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#### Abstract

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## 2 Cod in subareas 1 and 2 (Norwegian coastal waters)

## Gadus morhua - cod.27.1-2coastN and cod.27.2.coastS

A benchmark assessment (WKBARFAR) was conducted in February 2021 in order to address the failure of the current management plan to reduce fishing mortality on Norwegian coastal cod (ICES 2021a). The main outcome of the benchmark was that from assessment year 2021 onwards, Norwegian coastal cod (NCC; formally cod.27.1-2coast) will be split into two stocks/components by 67 degrees latitude-a data-rich one in the north: cod.27.1-2coastN (northern Norwegian coastal cod); and a data-limited one in the south: cod.27.2coastS (southern Norwegian coastal cod; Figure 2.0.1). The majority (approximately $80-90 \%$ ) of NCC catches are taken north of $67^{\circ} \mathrm{N}$ (Table 2.1.1), and this is also where the coastal survey has the best coverage. Genetic studies have revealed a genetic gradient in cod along the Norwegian coast without areas of distinct breaks in population connectivity (Dahle et al., 2018). However, NCC in northern Norway have more genetic material in common with the Northeast Arctic cod (NEAC; cod.27.1-2), compared to Norwegian coastal cod further south (Dahle et al., 2018).
Recent updates of the catch series, a revision of the acoustic survey index and a new swept-area index have improved the data basis for assessment in the northern area. The data for northern Norwegian coastal cod were considered of high enough quality to support an age-based analytical assessment. Southern Norwegian coastal cod $\left(62-67^{\circ} \mathrm{N}\right)$ represents the remaining commercial catches of NCC north of $62^{\circ} \mathrm{N}$ (approximately $10-20 \%$ ) and is not as consistently covered by the main survey relevant to monitoring cod. Current data availability and quality cannot support a full analytical assessment, and a data-limited approach has therefore been developed to support management of this stock.



Figure 2.0.1 Norwegian catch reporting areas used to define stock distribution areas for northern Norwegian coastal cod (left) and southern Norwegian coastal cod (right).

### 2.1 Fisheries (both stocks)

Coastal cod is fished throughout the year and within nearly all the distribution areas in the Norwegian statistical areas $03,04,05,00,06,07$ (Figure 2.0.1). Most of the coastal cod catches are taken as a bycatch in fisheries aimed at Northeast Arctic cod during its spawning and feeding migrations to coastal waters. The main fishery for coastal cod, therefore, takes place in the first half of the year. The main fishing areas are along the coast from Varangerfjord to Lofoten (areas $03,04,05,00$ ).

Recreational and tourist fisheries take an important fraction of the total catches in some local areas, especially near the coastal cities, and in some fjords where commercial fishing activity is low. However, there are a few reports trying to assess the amount in certain years. In 2010, these reports were used to construct a time-series of recreational catches (ICES 2010). These catch estimates are quite uncertain. No additional information was included during 2010-2018, and the annual recreational catch during this period has been assumed equal to the one estimated for 2009 (12 700 t).

A new project was conducted in the period 2017-2020 by IMR in collaboration with several Norwegian institutions (NINA, Akvaplan-niva, NMBU and Nordland Research), and a number of international partners. Three study areas Troms, Hordaland, and Oslofjord, were chosen because they represent contrasts in recreational fishing. The project is currently being finished and reports will follow, but some preliminary results were presented at the benchmark assessment (WKBARFAR WD13, ICES 2021a), and further used in the present coastal cod assessments.

Historically there has been no reporting system for NCC taken by recreational or tourist fishers in Norway. In 2019, the Norwegian Directorate for Fisheries established a web portal for obligatory catch reporting (both kept and released fish) by all registered fishing businesses. Tourist fishing effort related to tourist fishing businesses has about doubled from 2009 to 2019. The total quantity of cod caught by tourists staying in tourist businesses has also more than doubled from 1586 tonnes in 2009 (Vølstad et al., 2011) to about 3455 tonnes in 2019.

The current (2019) documented estimate of about 9000 tonnes (WKBARFAR WD13, ICES 2021a) is clearly an underestimate as tourists outside registered tourist businesses and residents fishing with fixed gears are not included. In the estimate of 9000 tonnes is also a share of the catch taken by anglers and released again. Based on investigations in other countries, the AFWG anticipates a mortality rate of $100 \%$ of fish caught by rod from land, and $20 \%$ of released cod caught by rod and handline at sea (e.g. Weltersbach and Strehlow, 2013; Capizzano et al., 2016). Until there is a better quantification of the missing recreational segments, the benchmark WK proposed to keep the quantity of 12700 tonnes recreational catch of Norwegian coastal cod north of $62^{\circ} \mathrm{N}$ on top of the commercial reported landings, with 7900 tonnes north of $67^{\circ} \mathrm{N}$ and 4800 tonnes between $62-67^{\circ} \mathrm{N}$ (Table 2.1).

The catch reporting (both kept and released fish) by the registered fishing businesses to the Norwegian Directorate of Fisheries in the corona-year 2020 shows a $77 \%$ decrease in catches of NCC compared to 2019. In the current assessment, the WG has taken this into account and reduced the rod and line catches from boats accordingly and kept the other recreational catches unchanged compared to 2019. This results in total 10039 tonnes unreported NCC caught by recreational fishers north of $62^{\circ} \mathrm{N}$ in 2020 , with 6233 tonnes caught north of $67^{\circ} \mathrm{N}$ and 3806 tonnes between $62-67^{\circ} \mathrm{N}$.

The total catch numbers-at-age (Tables 2.2.3c and Table 2.3.3) have been upscaled from the estimated catch-at-age in the commercial landings, according to the added amount in tonnes.

It is necessary to update the recreational catch with a better estimate as soon as this is available.

### 2.1.1 Revision of catch data

The benchmark assessment (WKBARFAR, ICES 2021a) tested and analysed two major catch data revisions: i) using the ECA model to separate the Norwegian coastal cod and the Northeast Arctic cod in the commercial catches by the structure of the otoliths in commercial samples, and ii) revising the catch in tonnes since 1992 using recommended seasonal product-round fish conversion factors instead of fixed factors for the whole year.

Until 1992, Norway used seasonal conversion factors to convert the weight of "headed-and-gutted" cod to round weight ( 1.6 during winter and 1.4 during the rest of the year). From 1992 onwards, this factor was set to 1.50 for the same product in all Norwegian cod fisheries all year around. From 2000 onwards, this factor was also agreed upon by the Joint Norwegian-Russian Fisheries Commission (JNRFC). From 2000, it hence became constant for all cod fisheries at all times of the year, although there is a larger difference between "headed-and-gutted" weight and round weight in the winter season when at least the Norwegian coastal fisheries for cod are dominated by mature fish with gonads.

Based on a report published by the Norwegian Directorate of Fisheries in 2015 (Blom, 2015), and summaries of this previously reported to the AFWG as WD 15 in 2017 and as WD 09 in 2020 (Nedreaas, 2017; Fotland and Nedreaas, 2020), ICES advice for NEA cod in 2018 states that "The use of constant conversion factors between round and gutted weight for all seasons and areas introduces a bias to the catch statistics". During the benchmark meeting (WKBARFAR, ICES 2021a) the Norwegian landings of cod by vessels below 28 m in January-April, all gears, were hence corrected by using 1.311 and 1.671 for the products "gutted with head" and "gutted without head", respectively, for each year since 1994.
Catch numbers-at-age are estimated for both stocks of NCC (i.e. northern and southern) by the ECA model. The commercial catches have been calculated back to 1984, but for the current assessment revised catch data were available for the period 1994-2020 for both stocks. The plan is to revise the catch data for both NCC stocks back to 1984.

### 2.1.2 Catch sampling

The basis for estimating Norwegian coastal cod catches is the total landings of cod from fisheries operating within the Norwegian statistical areas $03,04,05,00,06,07$ (ref. Figure 2.0.1), combined with the catch samplings of these fisheries. Commercial catches of cod are separated into types of cod by the structure of the otoliths in the commercial catch samples. Figure 2.1.2 illustrates the main difference between the two types: The figure and the following text is from (Berg et al., 2005):

Coastal cod has a smaller and more circular first translucent zone than northeast Arctic cod, and the distance between the first and the second translucent zone is larger. The shape of the first translucent zone in northeast Arctic cod is similar to the outer edge of the broken otolith and to the subsequent established translucent zones. This pattern is established at an age of 2 years, and error in differentiating between the two major types does not increase with age since the established growth zones do not change with age.

The precision and accuracy of the separation method for categorizing cod-type was investigated by comparing the results of different otolith reads to the results of genetic analyses, and the investigation determined that the results from the otolith method are high in accuracy (Berg et al., 2005). Nevertheless, in cases with a low percentage misclassification of large catches of pure NEA cod, the catches of coastal cod could be severely overestimated.


Figure 2.1.2. An image of a Norwegian coastal cod otolith (top) and a Northeast Arctic cod otolith (bottom). The two first translucent zones are highlighted. (from Berg et al., 2005).

Since the catches are separated by type of cod by the structure of the otoliths, the numbers of age samples are critical for the estimated catch of coastal cod. Table 2.1.2 shows the sampling of the cod fisheries by quarters, split by NCC and NEAC. The Norwegian sampling program changed in 2010, which led to poor sampling in that year. The sampling in later years gradually improved, and the number of samples (but not the number of otoliths) is now well above the level prior to 2010.

The number of otoliths sampled in 2020 is lower than in 2018 and 2019 due to reduced access to fish landing sites because of COVID-19, but the proportion of NCC in samples was similar; a total of 9012 fish were aged in 2020, whereof $37 \%$ were classified as Norwegian coastal cod.

### 2.1.3 Regulations

The Norwegian cod TAC is a combined TAC for both the NEAC stock and NCC stocks. Landings of cod are counted against the overall cod TAC for Norway, where the expected catch of NCC (North and South) is in the order of $10 \%$. The NCC part of this combined quota was set 40000 t in 2003 and earlier years. In 2004, it was set to 20000 t , and in the following years to 21000 t . There are no separate quotas given for the coastal cod for the different groups within the fishing fleet. Catches of coastal cod are thereby not effectively restricted by quotas.

Since the coastal cod is fished under a merged Norwegian coastal cod/Northeast Arctic cod quota, the main objective of these regulations is to move the traditional coastal fishery from areas with high fractions of NCC to areas where the proportion of NEAC is higher. Most regulation measures for NEAC also applies to NCC; minimum catch size, minimum mesh size, maximum bycatch of undersized fish, closure of areas having high densities of juveniles, and some seasonal and area restrictions. A number of regulations contribute to some protection of NCC, e.g. a ban
on trawl fishing inside 6 nautical miles from the baseline and "fjord-lines" that were drawn along the coast to close the fjords for direct cod fishing with vessels larger than 15 metres. For more details about the technical regulations, see ICES (2020).

Table 2.1.1. Left: estimated commercial catches of Norwegian coastal cod North of $67^{\circ} \mathrm{N}$ (NCC North) and between 62$67^{\circ} \mathrm{N}$ (NCC South), and Northeast Arctic cod between 62-67 ${ }^{\circ}$ (NEAC South). Middle: estimated recreational catches of cod north of $67^{\circ} \mathrm{N}$ and between $62-67^{\circ} \mathrm{N}$, all assumed to be coastal cod. Right: Recreational catches of NCC North and South that were sold and included in the commercial catch statistics. Note that an initial unlikely low share of NCC vs. NEAC in the $\mathbf{2 0 0 1}$ commercial landings compared to years before/after was replaced by an average of the 2000 and 2002 NCC values.

|  | Commercial catch (tonnes): |  |  | Recreational catch (tonnes): |  |  | Sold recreational catch included in commercial catch (tonnes)*: |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NCC North | NCC South | NEAC <br> South | NCC North | NCC South | Total | NCC North | NCC South | Total |
| 1994 | 52579 | 6381 | 23430 | 9144 | 5556 | 14700 |  |  |  |
| 1995 | 56907 | 8936 | 16981 | 9144 | 5556 | 14700 |  |  |  |
| 1996 | 41820 | 6207 | 13250 | 9020 | 5480 | 14500 |  |  |  |
| 1997 | 46605 | 4746 | 12695 | 9020 | 5480 | 14500 |  |  |  |
| 1998 | 45462 | 6200 | 9389 | 9082 | 5518 | 14600 |  |  |  |
| 1999 | 38743 | 5522 | 7101 | 8646 | 5254 | 13900 |  |  |  |
| 2000 | 33081 | 5838 | 4329 | 8460 | 5140 | 13600 |  |  |  |
| 2001 | 24470 | 5250 | 3499 | 8335 | 5065 | 13400 |  |  |  |
| 2002 | 32188 | 6937 | 4266 | 8460 | 5140 | 13600 |  |  |  |
| 2003 | 29253 | 8905 | 3943 | 8646 | 5254 | 13900 |  |  |  |
| 2004 | 31198 | 6866 | 3941 | 8335 | 5065 | 13400 |  |  |  |
| 2005 | 30097 | 8005 | 1462 | 8211 | 4989 | 13200 |  |  |  |
| 2006 | 36884 | 8612 | 1175 | 8087 | 4913 | 13000 |  |  |  |
| 2007 | 26200 | 7695 | 2250 | 8087 | 4913 | 13000 |  |  |  |
| 2008 | 27711 | 9889 | 1376 | 7962 | 4838 | 12800 |  |  |  |
| 2009 | 22988 | 7145 | 2474 | 7900 | 4800 | 12700 |  |  |  |
| 2010 | 34804 | 7634 | 2685 | 7900 | 4800 | 12700 |  |  |  |
| 2011 | 27982 | 7128 | 7474 | 7900 | 4800 | 12700 |  |  |  |
| 2012 | 26778 | 8187 | 4942 | 7900 | 4800 | 12700 | 1425 | 239 | 1665 |
| 2013 | 21376 | 5131 | 8395 | 7900 | 4800 | 12700 | 450 | 167 | 617 |
| 2014 | 22750 | 6244 | 6682 | 7900 | 4800 | 12700 | 774 | 229 | 1003 |
| 2015 | 34483 | 5004 | 5424 | 7900 | 4800 | 12700 | 618 | 226 | 844 |


| Commercial catch (tonnes): | Recreational catch (tonnes): | Sold recreational catch included in <br> commercial catch (tonnes)*: |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | NCC North | NCC South | NEAC <br> South | NCC North | NCC South | Total | NCC North | NCC South | Total |
| 2016 | 49503 | 5962 | 2006 | 7900 | 4800 | 12700 | 810 | 332 | 1142 |
| 2017 | 54273 | 4159 | 1242 | 7900 | 4800 | 12700 | 772 | 307 | 1078 |
| 2018 | 34532 | 4436 | 1822 | 7900 | 4800 | 12700 | 1206 | 340 | 1546 |
| 2019 | 35861 | 2965 | 1677 | 7900 | 4800 | 12700 | 1603 | 339 | 1943 |
| 2020 | 43133 | 3481 | 987 | 6233 | 3806 | 10039 | 1785 | 347 | 2132 |

*Source: Norwegian Directorate of Fisheries. All reported recreational cod assumed to be coastal cod.

Table 2.1.2. Number of otoliths sampled by quarter from commercial catches. NCC: Norwegian coastal cod. NEAC: Northeast Arctic cod. The table includes all otoliths from the Norwegian catch sampling areas $\mathbf{0}$ and $\mathbf{3 - 7}$ (covering both Norwegian coastal cod stocks).

| Year | Quarter 1 |  | Quarter 2 |  | Quarter 3 |  | Quarter 4 |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NCC | NEAC | NCC | NEAC | NCC | NEAC | NCC | NEAC | NCC | NEAC | \%NCC |
| 1985 | 1451 | 3852 | 777 | 1540 | 1277 | 1767 | 1966 | 730 | 5471 | 7889 | 41 |
| 1986 | 940 | 1594 | 1656 | 2579 | 0 | 0 | 669 | 966 | 3265 | 5139 | 39 |
| 1987 | 1195 | 2322 | 937 | 3051 | 638 | 1108 | 1122 | 1137 | 3892 | 7618 | 34 |
| 1988 | 257 | 546 | 160 | 619 | 87 | 135 | 55 | 44 | 559 | 1344 | 29 |
| 1989 | 556 | 1387 | 72 | 374 | 65 | 501 | 97 | 663 | 790 | 2925 | 21 |
| 1990 | 731 | 2974 | 61 | 689 | 252 | 97 | 265 | 674 | 1309 | 4434 | 23 |
| 1991 | 285 | 1168 | 92 | 561 | 77 | 96 | 279 | 718 | 733 | 2543 | 22 |
| 1992 | 152 | 619 | 281 | 788 | 79 | 82 | 272 | 672 | 784 | 2161 | 27 |
| 1993 | 314 | 1098 | 172 | 1046 | 0 | 0 | 310 | 541 | 796 | 2685 | 23 |
| 1994 | 317 | 1605 | 179 | 923 | 21 | 31 | 126 | 674 | 643 | 3233 | 17 |
| 1995 | 188 | 1591 | 232 | 1682 | 2095 | 1057 | 752 | 1330 | 3267 | 5660 | 37 |
| 1996 | 861 | 5486 | 591 | 1958 | 1784 | 1076 | 958 | 2256 | 4194 | 10776 | 28 |
| 1997 | 1106 | 5429 | 367 | 2494 | 1940 | 894 | 1690 | 1755 | 5103 | 10572 | 33 |
| 1998 | 608 | 4930 | 552 | 1342 | 489 | 1094 | 2999 | 2217 | 4648 | 9583 | 33 |
| 1999 | 1277 | 4702 | 493 | 2379 | 202 | 717 | 961 | 1987 | 2933 | 9785 | 23 |
| 2000 | 1283 | 4918 | 365 | 2112 | 386 | 1295 | 472 | 668 | 2506 | 9993 | 20 |
| 2001 | 1102 | 5091 | 352 | 2295 | 126 | 786 | 432 | 983 | 2012 | 9155 | 18 |
| 2002 | 823 | 5818 | 321 | 1656 | 503 | 831 | 897 | 1355 | 2544 | 9660 | 21 |


| Year | Quarter 1 |  | Quarter 2 |  | Quarter 3 |  | Quarter 4 |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NCC | NEAC | NCC | NEAC | NCC | NEAC | NCC | NEAC | NCC | NEAC | \%NCC |
| 2003 | 821 | 4197 | 445 | 2850 | 790 | 936 | 1112 | 1286 | 3168 | 9269 | 25 |
| 2004 | 1511 | 7539 | 758 | 2565 | 532 | 685 | 531 | 1317 | 3332 | 12106 | 22 |
| 2005 | 1583 | 6219 | 767 | 4383 | 473 | 258 | 877 | 1258 | 3700 | 12188 | 23 |
| 2006 | 2244 | 5087 | 1329 | 2819 | 590 | 271 | 119 | 71 | 4282 | 8248 | 34 |
| 2007 | 1867 | 5895 | 944 | 2496 | 503 | 648 | 637 | 1163 | 3951 | 10202 | 28 |
| 2008 | 1450 | 4162 | 1116 | 3122 | 626 | 515 | 693 | 999 | 3885 | 8798 | 31 |
| 2009 | 1114 | 5109 | 558 | 2592 | 126 | 253 | 842 | 465 | 2640 | 8419 | 24 |
| 2010 | 736 | 2000 | 572 | 992 | 464 | 195 | 325 | 270 | 2097 | 3457 | 38 |
| 2011 | 643 | 2271 | 789 | 2548 | 412 | 296 | 732 | 443 | 2576 | 5558 | 32 |
| 2012 | 1294 | 6283 | 749 | 1864 | 379 | 85 | 324 | 185 | 2746 | 8417 | 25 |
| 2013 | 966 | 5389 | 832 | 3155 | 216 | 88 | 1115 | 385 | 3129 | 9017 | 26 |
| 2014 | 1019 | 4470 | 869 | 3312 | 338 | 29 | 1060 | 524 | 3286 | 8335 | 28 |
| 2015 | 746 | 7770 | 618 | 3619 | 327 | 354 | 511 | 547 | 2202 | 12290 | 15 |
| 2016 | 2465 | 5581 | 1073 | 2445 | 616 | 207 | 1501 | 727 | 5655 | 8960 | 39 |
| 2017 | 2276 | 4568 | 879 | 2742 | 810 | 151 | 1231 | 475 | 5196 | 7936 | 40 |
| 2018 | 2007 | 4927 | 924 | 1882 | 498 | 104 | 1143 | 435 | 4572 | 7348 | 40 |
| 2019 | 1830 | 4594 | 759 | 1969 | 838 | 260 | 1284 | 445 | 4711 | 7268 | 39 |
| 2020 | 1926 | 3551 | 587 | 1688 | 424 | 85 | 434 | 317 | 3371 | 5641 | 37 |
| Av85-20 | 1110 | 4021 | 617 | 2087 | 527 | 472 | 800 | 852 | 3054 | 7461 | 29 |

### 2.2 Cod in subareas 1 and 2, north of $67^{\circ} \mathrm{N}$ (northern Norwegian coastal cod)

### 2.2.1 Stock status summary

An assessment based on the decisions of the 2021 WKBARFAR benchmark (ICES 2021a) is presented for this stock.

The 2021 assessment shows that SSB declined from a level just above Blim at the start of the assessment period (1994) to a low level in 1999. Between 1999-2002, SSB increased, but to a level lower than the one observed at the start of the assessment period. After 2002, SSB stayed at a similar level until 2010, after which it increased to approximately 50000 t lower than the 1994 level. After 2016, there has been a declining trend back towards the level estimated in 2003-2010, followed by an increase from 2019 to 2020 of approximately 10000 t . Fishing mortality mainly follows the trend in SSB, with highest F in the period with lowest estimated SSB. However, F
was higher at the start of the assessment period compared to 2013-2014, although SSB was higher in the first period. F also increased from 2019 to 2020 despite increasing SSB. Recruitment peaked in 1996 and has not been as high since. Comparatively good recruitment was seen in 2013-2018, after which it declined in 2019. In 2020, recruitment was the lowest observed since 2006, and the third-lowest observed in the time-series. TSB in 2020 is 9500 t lower than in 2019 and the lowest observed since 2013.

No previous advice has been issued for this stock. The 2021 advice for the previous Norwegian coastal cod stock (comprising the two new stocks) was to follow the Norwegian management plan, which implied reducing fishing mortality to 0.1 .

Further details on the stock assessment procedure can be found in the Stock Annex.

### 2.2.2 The fishery (Table 2.2.1-Table 2.2.4)

Commercial landings of northern Norwegian coastal cod in 2020 were 43133 t . Of the total landings, $28 \%$ were taken in ICES Division 1.b and the rest in Division 2.a (Table 2.2.1). The highest landings were made in the Norwegian catch reporting areas 03 and 04 , using Danish seine, longline and jig (Table 2.2.2). In total, a third of the landings were taken in gillnet fisheries, while trawl made up approximately $12 \%$ of landings.

The level of discarding and misreporting from coastal vessels has been investigated for three periods: 2000 and 2002-2003 (WD 14 at 2002 WG), and 2012-2018 (Berg and Nedreaas 2021). The report from the 2000-investigation concluded that there was both discarding and misreporting by species in 2000. In the gillnet fishery for cod, discarding and misreporting represented approximately $8-10 \%$ relative to reported catch, and $1 / 3$ of this was probably coastal cod. Data from 2002-2003 showed that misreporting in the coastal gillnet fisheries had been reduced significantly since 2000. A recent work by Berg and Nedreaas (2021) estimating discards of cod in the coastal gillnet fisheries during 2012-2018 showed that discarding (as percentage of total catch in weight including discards) decreased from less than $1 \%$ at the beginning of the period to less than $0.5 \%$ during 2016-2018. In weight, this corresponds to a decrease from more than 500 tonnes-per-year to about 180 tonnes-per-year. The reason for discarding seems to be highgrading by size (and price) during the first half of the year, and damaged fish (same size as landed fish) in the second half of the year.

Tourist fishing businesses reporting to the Norwegian Directorate of Fisheries in 2019 showed that about $42 \%$ of the reported rod and line catch was released, and with an assumed mortality of $20 \%$ of the released cod from the boat (see section 2.1 ), this corresponds to about $8 \%$ discards (dead fish) in the rod and line sector of the recreational fishery.

In the stock assessment, discarding is not included in the commercial landings, i.e. commercial catches are assumed equal to landings, but discarding in the rod and line (from boat) sector of the recreational fishery is included in the recreational catch estimate.

### 2.2.3 Survey results

A trawl-acoustic survey along the Norwegian coast from the Russian border to $62^{\circ} \mathrm{N}$ was started in autumn 1995. In 2003, this survey was combined with the former saithe survey at the coastal banks and moved from September to October-November (ICES acronym: A6335). Since then, the survey design included fixed bottom trawl stations in addition to trawl hauls set out on acoustic registrations. The seabed along the Norwegian coast is rugged, with sharp drops and peaks over short distances. This makes it difficult to get reliable survey indices both with acoustics and bottom trawl sampling. Acoustics can reach areas where the seabed is too uneven to perform bottom trawling, but species detection and discrimination can be hindered by dead
zones and acoustic shadows. Acoustics and bottom trawl data therefore contain both independent and overlapping information. For the 2021 benchmark, one acoustic and one swept-area index was prepared (WD 06 to AFWG 2021), and it was decided to include them both in the assessment. It should be noted that the uncertainties associated with the indices are rather large and increasing with age.

The survey indices are calculated with the software StoX (Johnsen et al., 2019), developed at the Institute of Marine Research in Norway. Instead of conventional age-length keys, StoX uses an imputation algorithm to assign age information to individuals that have been length measured but not aged. Crucial to coastal cod, the software also imputes other biological information, particularly otolith type, which is used to split the index on NEAC and NCC. The underlying assumption is that the proportion of NCC in length samples are representative of the proportion in the environment. StoX also estimates coefficients of variation using a bootstrap routine. The bootstrapping consists of two parts; resampling of primary sampling units (trawl stations or acoustic transects) with replacement, and the imputation of missing ages by random draw from individuals in the same length group. Primarily, age information is drawn from individuals in the same length group sampled in the same trawl haul. Should there be none, the draw extends to all trawl hauls within the same survey strata, and lastly, to the entire survey area. The CV is the variability resulting from both parts of the bootstrap routine.

The results of the 2020 survey (Staby et al., 2021) north of $67^{\circ} \mathrm{N}$ are presented in Tables 2.2.52.2.12.

### 2.2.3.1 Indices of abundance and survey mortality (Tables 2.2.5-2.2.8, Figures 2.2.2-2.2.4)

Both the acoustic (Table 2.2.5) and swept-area (Table 2.2.7) survey indices are lower in 2020 than in 2019, for nearly all age groups. The 2020 estimates of age 1 and 2 abundance are particularly low. The coefficient of variation (CV) is generally higher for ages 8 and above where there is less data. Both acoustic and swept-area index CVs for age 1, 9, and 10 were higher in 2020 than in 2018 and 2019, reflecting the low abundances of these age groups (Tables 2.2.6 and 2.2.8).

Survey mortality increased in 2020 relative to 2019, for most age groups (Figure 2.2.4). Generally, internal consistencies are low in both survey indices, and consequently, the survey mortality is highly variable between years (Figure 2.2.4).

### 2.2.3.2 Age reading and stock separation (Table 2.2.9)

About 2500 cod otoliths were sampled north of $67^{\circ} \mathrm{N}$ during the 2020 survey, which is up from 2100 in 2019 and the largest number of samples since 2003 (Table 2.2.9). The proportions of NCC at age among those otoliths were similar to previous years (Table 2.2.9). An error was discovered in the separation of stocks after AFWG was conducted. This error resulted in too few fish being categorized as coastal cod in 2020, and hence an erroneously low value for the coastal cod survey index in 2020. This error only affects northern coastal cod, and only in 2020. The error has been corrected, and the data and results presented here are based on the corrected data.

### 2.2.3.3 Length and weights-at-age (Tables 2.2.10-2.2.11, Figure 2.2.5)

Mean lengths-at-age in 2020 were similar to previous years (Table 2.2.10). Mean weight at age 1 was higher than in 2019, while it was similar for the other ages (Table 2.2.11). For age 8 and older the mean lengths and weights show larger variations, probably caused by few fish sampled in some years (Figure 2.2.5).

### 2.2.3.4 Maturity-at-age (Table 2.2.12, Figure 2.2.6)

The fraction of mature fish in the autumn survey (Table 2.2.12) show rather large variation between years. While some of the variation is likely related to variation in stock size and size at
age, it may also be partly caused by the difficulty of distinguishing mature and immature cod in autumn. Coastal cod spawn in February-June and many mature individuals are therefore in a resting state at the time of the survey in October-November. As part of the 2021 benchmark, the maturity ogive was recalculated to include spent/resting individuals to address this discrepancy. This gave an ogive similar to that estimated from a smaller fishery-dependent dataset, collected during the spawning season. In 2020, the proportion mature at age $2-7$ increased relative to 2020, while it decreased for age 8 (Figure 2.2.6). The proportion mature at age 2 in 2020 was particularly high, at a level not seen since 2008.

### 2.2.4 Data used in the Assessment

### 2.2.4.1 Catch numbers-at-age (Table 2.2.3c)

The estimated total catch-at-age (2-10+) for the period 1994-2020, including both commercial and recreational catches, is used in the assessment (Table 2.2.3c). Tables 2.2.3a and 2.2.3b show the commercial and recreational catches separately. The catch of ages $4-7$ were higher in 2020 than in the two previous years, while the catch of age $10+$ were about half compared to the two previous years. The total catch in tonnes increased by 5500 t compared to 2019.

### 2.2.4.2 Catch weight-at-age (Table 2.2.4)

Weight-at-age in catches is derived from the commercial sampling and is shown in Table 2.2.4. The same weight-at-age is assumed for recreational and tourist catches. Mean weights of ages 25 in 2020 are the highest observed in the time-series. Weight of the plus group is an average for the ages included in the plus group, weighted by abundance at age.

### 2.2.4.3 Tuning data (Table 2.2.13)

The acoustic and swept-area survey indices for ages $2-10+$ are used in the assessment (Table 2.2.13). The acoustic index is split in two parts; 1995-2002 and 2003-due to a change in catchability when fixed bottom trawl stations were introduced in the survey.

### 2.2.4.4 Stock weight-at-age (Table 2.2.14)

The weight-at-age for ages $2-7$ in the stock (Table 2.2.14) is obtained from the Norwegian coastal survey (Table 2.2.11), while catch weight-at-age (Table 2.2.4) is used for ages $8-10+$ due to large uncertainty for these ages in survey data (Figure 2.2.5). The survey weights are assumed to be relevant to the weight-at-age in the stock at survey time (October). These weights will, however, overestimate the stock biomass at the start of the year, and in the assessment model, SSB is therefore calculated after applying $80 \%$ of the year's fishing and natural mortality, corresponding to the survey timing.

### 2.2.4.5 Maturity-at-age (Table 2.2.12)

Annual maturity-at-age observed in the survey is used in the assessment (Table 2.2.12). Maturity of the plus group is an average for the ages included in the plus group, weighted by abundance-at-age.

### 2.2.4.6 $\quad$ Natural mortality (Table 2.2.15)

In Northeast Arctic cod, cannibalism has been documented to be a significant source of mortality that varies in relation to alternative food and in relation to the abundance of large cod. This might also be the case for the coastal cod (Pedersen and Pope 2003a and b). In the 2005 coastal cod survey 1125 cod stomachs were analysed (Mortensen 2007). The observed average frequency of occurrence of cod in cod stomachs was around $4 \%$. Other important predators on cod in coastal waters are cormorants, harbour porpoises and otters (Anfinsen 2002; Pedersen et al., 2007; Mortensen 2007). Young saithe (ages 2-4) has also been observed to consume post-larvae and 0-
group cod during summer/autumn (Aas 2007). As detailed data on consumption of coastal cod is lacking, natural mortality in the assessment is assumed dependent on cod size; M is calculated based on stock weight-at-age, following the method by Lorenzen (1996). With this method, M ranges from approximately 0.6 for age 2 to 0.2 for the plus group (Table 2.2.15).

### 2.2.5 Final assessment run

The 2021 assessment was run with the configuration decided upon at the 2021 benchmark (Table 2.2.16). The main features of the configuration are: 1 ) Coupling of fishing mortality states for ages $7-9,2$ ) Coupling of survey catchability parameters for ages 5-6 in the acoustic index part 1 and for ages 5-9 in the other two survey indices, 3) Separate variance parameter for age 2 in the catch, 4) $\operatorname{AR}(1)$-correlation between ages in the acoustic index part 2 and the swept-area index, and 5) Recruitment modelled as random walk.

The log-likelihood, number of parameters and AIC of the final run are presented in the table below. There were no problems with model convergence. In the 2021 assessment, there was no "base" (previous year's assessment) to compare with and the "Current" and "base" model are therefore the same.

| Model | Log(L) | \#par | AIC |
| :--- | :--- | :---: | :--- |
| Current | -180.17 | 37 | 434.33 |
| base | -180.17 | 37 | 434.33 |

The estimated survey catchabilities at age are presented in Table 2.2.17.

### 2.2.5.1 Model diagnostics (Figure 2.2.8-Figure 2.2.10)

A 5-year retrospective peel indicated no large problems with the estimates of SSB and Fbar (Figure 2.2.8). The second half of the model period has larger uncertainty as there is an additional survey index (from bottom trawl) that gives generally higher abundance estimates compared to the acoustic index. Mohn's rho (average 5-year retrospective bias) was 0.1 for SSB, -0.1 for Fbar, and 0.29 for recruitment. Thus, the model would have overestimated recruitment, particularly from 2013 and onwards, had it been run in previous years.

The process residuals were improved at the benchmark by splitting the acoustic index in two parts. Some clustering of positive/negative residuals remain in the $\log (\mathrm{N})$ residuals, with more negative residuals in the period 1995-2002 compared to the later period (Figure 2.2.9). The one-step-ahead residuals (Figure 2.2.10) were also improved by introducing correlations between ages in the survey indices. Evaluation of this correlation structure should be made at the next benchmark to see if the residuals can be further improved.

### 2.2.5.2 Model results (Table 2.2.18-2.2.20)

Recruitment in 2020 is the third-lowest estimate in the period covered by the model (Table 2.2.18). While SSB increased with 10000 t in 2020, Fbar also increased compared to 2019 reflecting an increase in catches of ages $4-7$ (Table 2.2.18 and Table 2.2.3c). Fishing mortality for ages 6-9 in 2020 were higher than in 2018 and 2019, while F for age 10+ was lower (Table 2.2.19). Abundances of ages 9 and 10+ in 2020 are the lowest seen since 2005 and 2009, respectively (Table 2.2.20). Abundance of ages 4 and 8 increased compared to 2019.

### 2.2.6 Reference points

Reference points were evaluated at the 2021 benchmark (ICES 2021a). The estimated stock-recruitment relationship showed increasing recruitment with increasing SSB throughout the model period, and the same pattern results from adding 2020 data in the assessment (Figure 2.2.11). At the benchmark, Blim was therefore set near the highest SSB observed, based on the reasoning that the lack of plateau in the SSB-recruit relationship indicates that the stock is below full reproductive capacity. In the assessment model, recruitment is at age 2. A similar pattern of increasing recruitment with SSB is evident when age 3 abundance is plotted against SSB (Figure 2.2.12).

No reference points for fishing mortality could be determined at the benchmark due to the lack of observations above Blim.

### 2.2.6.1 Management plan

No management plan is currently implemented for this stock.

### 2.2.7 Predictions

### 2.2.7.1 Input data (Tables 2.2.21a-b)

The built-in forecast option in SAM is used for short term prediction. Status quo fishing is assumed for the interim year, i.e. same F as in the final year of assessment (Table 2.2.21a). Process noise is included in the prediction (i.e. processNoiseF=FALSE). Averages from the last 5 years of the assessment are used for stock weights, catch weights, maturity, and natural mortality-at-age (Table 2.2.21b). Recruitment is the median resampled from the last 10 years (Table 2.2.21a).

### 2.2.7.2 Catch options for 2021 (Table 2.2.22, Figure 2.2.13)

The ICES advice basis for northern Norwegian coastal cod is the precautionary approach. This leads to catch advice of no more than 7865 tonnes in 2022. This catch level is expected to lead to a $25 \%$ increase in SSB relative to SSB estimated for 2021, while the same level of fishing in 2022 as in 2020 is expected to give a $0.15 \%$ decrease in SSB. Zero catch in 2022 is expected to give a $30 \%$ increase in SSB (Table 2.2.21, Figure 2.2.13).

### 2.2.7.3 Comparison of the present and last year's assessments

No previous assessment is available for this stock.

### 2.2.8 Comments to the assessment and the forecast

The assessment model performs rather well despite uncertainties in survey data. The main problem for this assessment is the lack of a full set of reference points and the uncertainty in the reference level for SSB. There is a need to perform further simulations to improve the reference points. Since this stock is part of a mixed fishery with Northeast Arctic cod and cannot be visually separated at sea, this year's catch advice is unlikely to be followed in practice. It is therefore advised to develop a management plan for this stock, detailing catch levels and regulations that may lead to the rebuilding of the stock over a longer period.

### 2.2.9 Tables and figures

Table 2.2.1. Northern Norwegian coastal cod. Total commercial catch (t) by fishing areas in 2020.

| Year | 03 | 04 | 00 | 05 | Total in Division 1.b <br> (NOR area 03) | Total in Division 2.a <br> (NOR areas 04+00+05) | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2020 | 12245 | 12393 | 7652 | 10832 | 12245 | 30877 | $43122^{*}$ |

*Differs slightly from Table 2.2.3a due to different spatial units used in estimation.

Table 2.2.2. Commercial catch of northern Norwegian coastal cod ( t ) in $\mathbf{2 0 2 0}$ by gear and Norwegian statistical fishing area.

| Year | 2020 | 04 | 00 | 05 | Total north of $67^{\circ} \mathrm{N}$ | \% by gear |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Area | 03 | 1259 | 3931 | 4018 | 3813 | 13021 |
| Gillnet |  |  |  |  | 30.2 |  |
| L.line/Jig | 1519 | 2342 | 0.2 | 1443 | 5304 | 12.3 |
| Danish seine | 9467 | 6120 | 3634 | 5576 | 24797 | 57.5 |
| Trawl | 12245 | 12393 | 7652 | 10832 | $43122^{* *}$ |  |
| Others* |  |  |  |  |  |  |
| Total |  |  |  |  |  |  |

*in 2020, longline, jig and Danish seine are all included in the 'others' category.
**Differs slightly from Table 2.2.3a due to different spatial units used in estimation.

Table 2.2.3a. Northern Norwegian coastal cod. Estimated commercial landings in numbers ('000) at-age and total tonnes by year.

|  | Age |  | $\mathbf{4}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1994 | 11 | 98 | 978 | 4394 | 3760 | 2756 | 1119 | 304 | 675 | Landed |
| 1995 | 21 | 228 | 814 | 2743 | 4796 | 3164 | 1815 | 943 | 612 | 56907 |
| 1996 | 41 | 768 | 1415 | 2035 | 3130 | 3086 | 1210 | 542 | 584 | 41820 |
| 1997 | 57 | 1111 | 2106 | 1956 | 2344 | 2721 | 1856 | 565 | 746 | 46605 |
| 1998 | 436 | 1631 | 6433 | 4391 | 2784 | 835 | 779 | 377 | 393 | 45462 |
| 1999 | 79 | 912 | 3395 | 4938 | 2037 | 783 | 527 | 394 | 425 | 38743 |
| 2000 | 30 | 534 | 2549 | 3925 | 2240 | 826 | 376 | 112 | 273 | 33081 |
| 2001 | 10 | 330 | 1863 | 2242 | 1641 | 961 | 305 | 104 | 493 | 24470 |
| 2002 | 42 | 308 | 1551 | 2585 | 2391 | 1057 | 630 | 183 | 363 | 32188 |
| 2003 | 120 | 350 | 952 | 1859 | 2173 | 1206 | 582 | 308 | 252 | 29253 |


|  | Age |  |  |  |  |  |  |  |  | Tonnes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Landed |
| 2004 | 23 | 179 | 1067 | 1520 | 2189 | 1570 | 784 | 328 | 371 | 31198 |
| 2005 | 13 | 241 | 924 | 1984 | 2003 | 1463 | 716 | 255 | 345 | 30097 |
| 2006 | 23 | 222 | 1276 | 1977 | 2619 | 1735 | 1017 | 402 | 396 | 36884 |
| 2007 | 36 | 376 | 1198 | 1667 | 1327 | 1088 | 477 | 277 | 279 | 26200 |
| 2008 | 63 | 387 | 997 | 1909 | 1549 | 1005 | 576 | 278 | 287 | 27711 |
| 2009 | 21 | 456 | 667 | 1177 | 1194 | 812 | 419 | 431 | 211 | 22988 |
| 2010 | 29 | 530 | 754 | 2832 | 1947 | 1055 | 528 | 283 | 857 | 34804 |
| 2011 | 65 | 465 | 1209 | 1318 | 1239 | 1081 | 568 | 343 | 583 | 27982 |
| 2012 | 374 | 1017 | 1126 | 1118 | 1287 | 760 | 364 | 177 | 596 | 26778 |
| 2013 | 131 | 503 | 1024 | 1038 | 909 | 704 | 478 | 219 | 340 | 21376 |
| 2014 | 88 | 505 | 824 | 1258 | 839 | 676 | 523 | 297 | 397 | 22750 |
| 2015 | 331 | 1106 | 1411 | 1251 | 1700 | 1040 | 639 | 437 | 873 | 34483 |
| 2016 | 75 | 937 | 1988 | 1582 | 1723 | 2119 | 1174 | 640 | 1073 | 49503 |
| 2017 | 846 | 1577 | 2071 | 2323 | 2087 | 1491 | 1331 | 700 | 903 | 54273 |
| 2018 | 171 | 563 | 1465 | 1634 | 1525 | 1416 | 747 | 518 | 497 | 34532 |
| 2019 | 49 | 953 | 1299 | 1776 | 1585 | 1260 | 985 | 318 | 519 | 35861 |
| 2020 | 40 | 534 | 2205 | 2116 | 2538 | 1615 | 906 | 354 | 309 | 43133 |

Table 2.2.3b. Northern Norwegian coastal cod. Estimated catch number ('000) at-age in recreational and tourist catches.

|  | Age |  | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year landed |  |  |  |  |  |  |  |  |  |  |
| 1994 | $\mathbf{2}$ | 17 | 170 | 764 | 654 | 479 | 195 | 53 | 117 | 9144 |
| 1995 | 3 | 37 | 131 | 441 | 771 | 508 | 292 | 151 | 98 | 9144 |
| 1996 | 9 | 166 | 305 | 439 | 675 | 666 | 261 | 117 | 126 | 9020 |
| 1997 | 11 | 215 | 408 | 378 | 454 | 527 | 359 | 109 | 144 | 9020 |
| 1998 | 87 | 326 | 1285 | 877 | 556 | 167 | 156 | 75 | 78 | 9082 |
| 1999 | 18 | 204 | 758 | 1102 | 455 | 175 | 118 | 88 | 95 | 8646 |
| 2000 | 8 | 136 | 652 | 1004 | 573 | 211 | 96 | 29 | 70 | 8460 |
| 2001 | 3 | 112 | 635 | 764 | 559 | 327 | 104 | 36 | 168 | 8335 |


| Year | Age |  |  |  |  |  |  |  |  | Tonnes <br> landed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |
| 2002 | 11 | 81 | 408 | 679 | 628 | 278 | 166 | 48 | 95 | 8460 |
| 2003 | 36 | 104 | 281 | 549 | 642 | 356 | 172 | 91 | 74 | 8646 |
| 2004 | 6 | 48 | 285 | 406 | 585 | 419 | 209 | 88 | 99 | 8335 |
| 2005 | 4 | 66 | 252 | 541 | 546 | 399 | 195 | 69 | 94 | 8211 |
| 2006 | 5 | 49 | 280 | 433 | 574 | 380 | 223 | 88 | 87 | 8087 |
| 2007 | 11 | 116 | 370 | 514 | 410 | 336 | 147 | 85 | 86 | 8087 |
| 2008 | 18 | 111 | 287 | 549 | 445 | 289 | 165 | 80 | 82 | 7962 |
| 2009 | 7 | 157 | 229 | 405 | 410 | 279 | 144 | 148 | 73 | 7900 |
| 2010 | 7 | 120 | 171 | 643 | 442 | 240 | 120 | 64 | 194 | 7900 |
| 2011 | 18 | 131 | 341 | 372 | 350 | 305 | 160 | 97 | 165 | 7900 |
| 2012 | 110 | 300 | 332 | 330 | 380 | 224 | 107 | 52 | 176 | 7900 |
| 2013 | 48 | 186 | 379 | 383 | 336 | 260 | 177 | 81 | 126 | 7900 |
| 2014 | 31 | 175 | 286 | 437 | 291 | 235 | 181 | 103 | 138 | 7900 |
| 2015 | 76 | 253 | 323 | 287 | 389 | 238 | 146 | 100 | 200 | 7900 |
| 2016 | 12 | 150 | 317 | 253 | 275 | 338 | 187 | 102 | 171 | 7900 |
| 2017 | 123 | 230 | 301 | 338 | 304 | 217 | 194 | 102 | 131 | 7900 |
| 2018 | 39 | 129 | 335 | 374 | 349 | 324 | 171 | 119 | 114 | 7900 |
| 2019 | 11 | 210 | 286 | 391 | 349 | 278 | 217 | 70 | 114 | 7900 |
| 2020 | 6 | 77 | 319 | 306 | 367 | 233 | 131 | 51 | 45 | 6233 |

Table 2.2.3c. Northern Norwegian coastal cod. Total estimated catch number ('000) at age, including recreational and tourist catches.

|  | Age |  |  | $\mathbf{4}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ | landed |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ |  |  |  |  |  | Tonnes |  |
| 1994 | 13 | 115 | 1148 | 5158 | 4414 | 3235 | 1313 | 356 | 793 | 61723 |
| 1995 | 24 | 264 | 945 | 3183 | 5567 | 3672 | 2106 | 1094 | 711 | 66051 |
| 1996 | 50 | 934 | 1720 | 2473 | 3805 | 3752 | 1471 | 659 | 709 | 50840 |
| 1997 | 68 | 1326 | 2514 | 2334 | 2797 | 3248 | 2215 | 674 | 890 | 55624 |
| 1998 | 523 | 1957 | 7718 | 5268 | 3341 | 1002 | 935 | 452 | 471 | 54544 |
| 1999 | 97 | 1116 | 4152 | 6040 | 2492 | 957 | 644 | 482 | 520 | 47390 |


|  | Age |  |  |  |  |  |  |  |  | Tonnes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | landed |
| 2000 | 38 | 670 | 3201 | 4929 | 2812 | 1037 | 472 | 141 | 342 | 41541 |
| 2001 | 13 | 442 | 2497 | 3006 | 2199 | 1288 | 409 | 140 | 661 | 32806 |
| 2002 | 53 | 389 | 1959 | 3265 | 3019 | 1335 | 796 | 231 | 459 | 40648 |
| 2003 | 156 | 454 | 1234 | 2408 | 2815 | 1562 | 754 | 399 | 326 | 37900 |
| 2004 | 30 | 227 | 1352 | 1926 | 2774 | 1989 | 993 | 415 | 470 | 39533 |
| 2005 | 17 | 307 | 1176 | 2525 | 2550 | 1862 | 911 | 324 | 440 | 38308 |
| 2006 | 28 | 271 | 1556 | 2410 | 3193 | 2115 | 1240 | 490 | 482 | 44970 |
| 2007 | 47 | 492 | 1567 | 2181 | 1737 | 1423 | 624 | 362 | 365 | 34287 |
| 2008 | 81 | 498 | 1284 | 2458 | 1994 | 1294 | 741 | 358 | 369 | 35674 |
| 2009 | 28 | 612 | 896 | 1582 | 1605 | 1091 | 563 | 579 | 284 | 30888 |
| 2010 | 35 | 651 | 925 | 3474 | 2388 | 1295 | 647 | 347 | 1051 | 42704 |
| 2011 | 83 | 597 | 1550 | 1690 | 1588 | 1386 | 728 | 440 | 747 | 35882 |
| 2012 | 484 | 1317 | 1458 | 1447 | 1666 | 984 | 471 | 229 | 772 | 34678 |
| 2013 | 179 | 689 | 1403 | 1421 | 1245 | 965 | 655 | 300 | 466 | 29276 |
| 2014 | 119 | 680 | 1110 | 1695 | 1130 | 911 | 704 | 400 | 534 | 30650 |
| 2015 | 407 | 1360 | 1734 | 1537 | 2089 | 1278 | 785 | 537 | 1072 | 42383 |
| 2016 | 86 | 1086 | 2305 | 1835 | 1998 | 2458 | 1362 | 743 | 1244 | 57403 |
| 2017 | 969 | 1806 | 2373 | 2661 | 2391 | 1707 | 1525 | 802 | 1035 | 62173 |
| 2018 | 210 | 691 | 1800 | 2007 | 1873 | 1740 | 918 | 637 | 611 | 42432 |
| 2019 | 60 | 1163 | 1585 | 2167 | 1934 | 1537 | 1202 | 387 | 633 | 43761 |
| 2020 | 45 | 612 | 2524 | 2422 | 2905 | 1849 | 1037 | 405 | 353 | 49366 |

Table 2.2.4. Northern Norwegian coastal cod. Mean catch weight at age (kg).

|  | Age |  | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ |  |  |  |  |  |
| 1994 | 0.910 | 1.422 | 1.987 | 2.649 | 3.479 | 4.343 | 5.245 | 6.487 | 8.825 |
| 1995 | 0.784 | 1.272 | 1.708 | 2.236 | 3.073 | 4.203 | 5.228 | 6.121 | 9.469 |
| 1996 | 0.874 | 1.269 | 1.722 | 2.385 | 2.968 | 3.660 | 4.544 | 5.462 | 7.814 |
| 1997 | 1.115 | 1.490 | 1.902 | 2.497 | 3.219 | 3.930 | 4.738 | 5.616 | 7.768 |


|  | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1998 | 0.719 | 1.212 | 1.654 | 2.343 | 3.346 | 3.969 | 4.786 | 5.389 | 9.584 |
| 1999 | 0.989 | 1.512 | 1.975 | 2.501 | 3.331 | 4.032 | 4.923 | 5.415 | 8.339 |
| 2000 | 1.019 | 1.452 | 2.057 | 2.598 | 3.447 | 4.449 | 5.553 | 5.834 | 9.781 |
| 2001 | 1.014 | 1.448 | 1.905 | 2.593 | 3.266 | 3.756 | 4.498 | 4.794 | 7.711 |
| 2002 | 0.929 | 1.470 | 2.059 | 2.760 | 3.590 | 4.467 | 5.268 | 6.236 | 9.943 |
| 2003 | 1.082 | 1.687 | 2.180 | 2.944 | 3.754 | 4.672 | 5.417 | 5.713 | 9.070 |
| 2004 | 1.145 | 1.604 | 2.186 | 2.848 | 3.640 | 4.555 | 5.367 | 5.930 | 7.991 |
| 2005 | 1.112 | 1.622 | 2.249 | 3.017 | 3.539 | 4.371 | 5.233 | 5.981 | 8.320 |
| 2006 | 1.522 | 2.020 | 2.491 | 3.284 | 4.075 | 4.887 | 5.806 | 6.638 | 9.710 |
| 2007 | 1.072 | 1.546 | 2.168 | 2.968 | 3.987 | 4.925 | 5.781 | 6.871 | 9.771 |
| 2008 | 1.153 | 1.663 | 2.355 | 3.043 | 3.970 | 4.902 | 5.844 | 6.279 | 9.239 |
| 2009 | 1.331 | 1.761 | 2.502 | 3.328 | 4.196 | 5.218 | 6.178 | 6.516 | 9.248 |
| 2010 | 1.252 | 1.770 | 2.375 | 3.103 | 3.834 | 4.483 | 5.437 | 6.185 | 7.599 |
| 2011 | 1.080 | 1.689 | 2.310 | 3.031 | 3.906 | 4.681 | 5.941 | 6.422 | 8.346 |
| 2012 | 1.010 | 1.653 | 2.328 | 3.232 | 4.246 | 5.111 | 6.448 | 6.914 | 9.446 |
| 2013 | 1.107 | 1.674 | 2.295 | 3.122 | 3.997 | 4.873 | 5.892 | 6.800 | 10.104 |
| 2014 | 1.187 | 1.788 | 2.410 | 3.222 | 4.118 | 5.165 | 5.791 | 6.461 | 9.643 |
| 2015 | 1.055 | 1.545 | 2.192 | 3.030 | 3.745 | 4.724 | 5.601 | 6.482 | 9.044 |
| 2016 | 1.279 | 1.774 | 2.363 | 3.171 | 3.972 | 4.868 | 5.893 | 6.850 | 8.928 |
| 2017 | 1.316 | 1.785 | 2.468 | 3.225 | 4.077 | 5.014 | 5.977 | 6.933 | 9.356 |
| 2018 | 1.141 | 1.700 | 2.307 | 3.090 | 3.878 | 4.770 | 5.711 | 6.581 | 9.333 |
| 2019 | 1.431 | 1.904 | 2.615 | 3.254 | 4.116 | 4.868 | 5.748 | 6.562 | 8.561 |
| 2020 | 1.487 | 2.147 | 2.823 | 3.514 | 4.218 | 4.932 | 5.655 | 6.387 | 9.024 |

Table 2.2.5. Northern Norwegian coastal cod. Acoustic abundance indices by age (in thousands) and total biomass ( $t$ ) from the Coastal survey (A6335). The split between coastal cod and Northeast Arctic cod is uncertain for age 1.

| Age |  |  |  | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ | Sum | Biomass |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ |  |  |  |  |  |  |  |  |
| 1995 | 26495 | 8774 | 4974 | 6382 | 6440 | 4373 | 1309 | 532 | 319 | 132 | 59729 | 55126 |
| 1996 | 17580 | 9025 | 8592 | 4576 | 5306 | 2723 | 1022 | 213 | 32 | 24 | 49093 | 39263 |


|  | Age |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Sum | Biomass |
| 1997 | 16567 | 15358 | 16930 | 7710 | 4484 | 2316 | 716 | 328 | 59 | 33 | 64502 | 45756 |
| 1998 | 8360 | 6757 | 8524 | 8261 | 3717 | 1530 | 700 | 102 | 122 | 45 | 38118 | 39474 |
| 1999 | 2494 | 3486 | 3387 | 2788 | 2498 | 751 | 172 | 30 | 22 | 20 | 15648 | 16167 |
| 2000 | 5028 | 7439 | 5831 | 3939 | 3853 | 2825 | 622 | 258 | 71 | 32 | 29899 | 35602 |
| 2001 | 2711 | 4551 | 4246 | 3776 | 2184 | 1499 | 974 | 149 | 29 | 93 | 20211 | 27250 |
| 2002 | 1188 | 2071 | 2532 | 2926 | 2075 | 970 | 596 | 293 | 106 | 124 | 12882 | 21203 |
| 2003 | 3276 | 2168 | 3026 | 3303 | 1838 | 1519 | 651 | 364 | 190 | 69 | 16403 | 23978 |
| 2004 | 3046 | 2643 | 2819 | 2589 | 1686 | 1094 | 371 | 213 | 104 | 72 | 14639 | 18237 |
| 2005 | 904 | 1201 | 2228 | 1816 | 1490 | 843 | 234 | 233 | 127 | 79 | 9156 | 14690 |
| 2006 | 4981 | 1836 | 2587 | 2210 | 1453 | 1612 | 1046 | 130 | 89 | 27 | 15970 | 22116 |
| 2007 | 2458 | 3037 | 2778 | 3794 | 2437 | 1632 | 1215 | 441 | 120 | 41 | 17952 | 33314 |
| 2008 | 2344 | 1739 | 1684 | 1511 | 985 | 761 | 399 | 225 | 97 | 74 | 9821 | 15491 |
| 2009 | 3907 | 1502 | 2084 | 2596 | 1373 | 605 | 386 | 378 | 140 | 64 | 13035 | 18716 |
| 2010 | 5509 | 2503 | 2853 | 2240 | 1679 | 583 | 309 | 432 | 229 | 195 | 16531 | 21966 |
| 2011 | 2104 | 2542 | 1869 | 2372 | 1469 | 1215 | 394 | 278 | 137 | 150 | 12529 | 23115 |
| 2012 | 3561 | 2170 | 3546 | 1832 | 1154 | 791 | 503 | 254 | 107 | 224 | 14142 | 20913 |
| 2013 | 4694 | 3084 | 1597 | 1770 | 1287 | 838 | 657 | 430 | 216 | 252 | 14825 | 21105 |
| 2014 | 6030 | 4171 | 3066 | 2137 | 2904 | 1609 | 1151 | 429 | 462 | 326 | 22286 | 37127 |
| 2015 | 3421 | 3122 | 2465 | 1802 | 1017 | 1128 | 477 | 363 | 303 | 265 | 14362 | 23144 |
| 2016 | 2921 | 3341 | 3667 | 2349 | 2308 | 841 | 669 | 452 | 222 | 308 | 17078 | 30763 |
| 2017 | 1018 | 3289 | 3202 | 2335 | 1764 | 1122 | 450 | 256 | 181 | 183 | 13800 | 25998 |
| 2018 | 4977 | 2847 | 1837 | 2376 | 1246 | 946 | 494 | 246 | 136 | 169 | 15274 | 22602 |
| 2019 | 2607 | 2992 | 3724 | 2221 | 2149 | 1272 | 656 | 212 | 262 | 266 | 16360 | 29992 |
| 2020 | 477 | 1619 | 3365 | 3564 | 1821 | 853 | 491 | 299 | 85 | 126 | 12702 | 25425 |

Table 2.2.6. Northern Norwegian coastal cod. Acoustic abundance index coefficient of variation (CV, in \%) by age.

|  | Age |  |  | $\mathbf{4}$ | $\mathbf{5}$ |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | 9 | 10 |
| 1995 | 17 | 13 | 9 | 12 | 14 | 21 | 19 | 40 | 51 | 41 |


|  | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1996 | 20 | 11 | 15 | 17 | 14 | 26 | 54 | 39 | 52 | 156 |
| 1997 | 24 | 25 | 16 | 16 | 14 | 25 | 26 | 47 | 90 | 81 |
| 1998 | 26 | 19 | 12 | 16 | 16 | 31 | 69 | 40 | 87 | 104 |
| 1999 | 24 | 10 | 11 | 20 | 17 | 23 | 19 | 47 | 40 | 92 |
| 2000 | 14 | 16 | 12 | 10 | 9 | 10 | 15 | 29 | 49 | 89 |
| 2001 | 18 | 31 | 18 | 16 | 19 | 18 | 21 | 41 | 72 | 69 |
| 2002 | 25 | 17 | 21 | 16 | 14 | 15 | 23 | 36 | 72 | 67 |
| 2003 | 27 | 26 | 14 | 14 | 14 | 16 | 18 | 22 | 26 | 35 |
| 2004 | 17 | 15 | 14 | 12 | 13 | 17 | 17 | 25 | 69 | 33 |
| 2005 | 18 | 23 | 18 | 10 | 14 | 20 | 23 | 30 | 40 | 61 |
| 2006 | 108 | 68 | 15 | 14 | 15 | 27 | 22 | 23 | 31 |  |
| 2007 | 21 | 20 | 19 | 15 | 16 | 16 | 21 | 31 | 45 | 97 |
| 2008 | 24 | 19 | 14 | 13 | 12 | 14 | 20 | 24 | 39 | 37 |
| 2009 | 22 | 20 | 15 | 12 | 17 | 14 | 18 | 19 | 31 | 25 |
| 2010 | 41 | 18 | 16 | 13 | 12 | 22 | 22 | 22 | 21 | 21 |
| 2011 | 22 | 17 | 16 | 15 | 15 | 15 | 27 | 21 | 19 | 35 |
| 2012 | 20 | 20 | 13 | 14 | 15 | 11 | 19 | 16 | 24 | 18 |
| 2013 | 14 | 16 | 14 | 15 | 14 | 13 | 17 | 20 | 31 | 37 |
| 2014 | 16 | 19 | 12 | 15 | 15 | 13 | 15 | 14 | 23 | 43 |
| 2015 | 21 | 16 | 11 | 10 | 12 | 12 | 16 | 16 | 16 | 27 |
| 2016 | 29 | 15 | 10 | 8 | 11 | 16 | 17 | 21 | 39 | 31 |
| 2017 | 34 | 16 | 12 | 16 | 14 | 18 | 23 | 28 | 43 | 25 |
| 2018 | 18 | 17 | 17 | 16 | 18 | 9 | 18 | 60 | 20 | 35 |
| 2019 | 18 | 20 | 15 | 13 | 12 | 15 | 18 | 28 | 33 | 35 |
| 2020 | 30 | 16 | 17 | 11 | 12 | 14 | 19 | 26 | 40 | 57 |

Table 2.2.7. Northern Norwegian coastal cod. Swept-area abundance indices by age (in thousands) and total biomass ( $t$ ) from the Coastal survey (A6335). The split between coastal cod and Northeast Arctic cod is uncertain for age 1.

| Age |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Sum | Biomass |
| 2003 | 5254 | 3268 | 3763 | 4521 | 2700 | 2319 | 863 | 489 | 220 | 69 | 23467 | 33861 |
| 2004 | 2837 | 2201 | 2396 | 2602 | 1463 | 722 | 359 | 181 | 46 | 63 | 12868 | 15980 |
| 2005 | 665 | 1042 | 1988 | 1478 | 1268 | 746 | 157 | 107 | 68 | 54 | 7574 | 11379 |
| 2006 | 1802 | 2156 | 2623 | 2946 | 1554 | 1026 | 941 | 171 | 107 | 23 | 13349 | 22526 |
| 2007 | 446 | 911 | 853 | 1071 | 789 | 465 | 394 | 114 | 75 | 29 | 5146 | 11943 |
| 2008 | 2463 | 1822 | 2795 | 1883 | 1419 | 1145 | 580 | 348 | 161 | 94 | 12710 | 23090 |
| 2009 | 6642 | 2251 | 3570 | 3716 | 1584 | 868 | 712 | 466 | 204 | 160 | 20172 | 24986 |
| 2010 | 7412 | 2353 | 3268 | 3385 | 2397 | 784 | 383 | 733 | 317 | 328 | 21360 | 29875 |
| 2011 | 2322 | 3471 | 2498 | 2866 | 2095 | 1445 | 292 | 315 | 213 | 310 | 15827 | 27845 |
| 2012 | 4299 | 3218 | 4485 | 2784 | 1537 | 1042 | 930 | 411 | 200 | 346 | 19251 | 28587 |
| 2013 | 6382 | 4101 | 1706 | 2666 | 1887 | 1575 | 890 | 578 | 297 | 419 | 20502 | 32875 |
| 2014 | 5696 | 5448 | 4026 | 3034 | 3521 | 2016 | 1388 | 465 | 364 | 337 | 26296 | 43823 |
| 2015 | 4298 | 4733 | 4154 | 3727 | 2068 | 1818 | 902 | 506 | 397 | 222 | 22827 | 40385 |
| 2016 | 3944 | 4433 | 4522 | 2610 | 1995 | 746 | 735 | 413 | 203 | 210 | 19810 | 31320 |
| 2017 | 768 | 2891 | 2407 | 1563 | 1151 | 715 | 308 | 200 | 147 | 157 | 10308 | 18682 |
| 2018 | 4070 | 3197 | 1916 | 1879 | 1049 | 748 | 323 | 183 | 128 | 168 | 13661 | 18815 |
| 2019 | 2234 | 2114 | 2470 | 1508 | 1460 | 839 | 490 | 148 | 129 | 211 | 11601 | 19974 |
| 2020 | 560 | 1670 | 2599 | 2416 | 1188 | 611 | 291 | 177 | 49 | 72 | 9632 | 14211 |

Table 2.2.8. Northern Norwegian coastal cod. Swept-area abundance index coefficient of variation (CV, in \%).

|  | Age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | 9 | 10 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2003 | 23 | 23 | 16 | 14 | 12 | 12 | 24 | 32 | 25 | 69 |
| 2004 | 27 | 16 | 16 | 16 | 21 | 21 | 23 | 34 | 40 | 37 |
| 2005 | 21 | 28 | 30 | 22 | 16 | 25 | 24 | 25 | 45 | 58 |
| 2006 | 20 | 34 | 24 | 26 | 17 | 13 | 24 | 30 | 34 |  |
| 2007 | 23 | 28 | 30 | 18 | 17 | 15 | 24 | 31 | 44 | 87 |
| 2008 | 15 | 26 | 21 | 13 | 11 | 17 | 15 | 20 | 37 | 36 |


| Year | Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | 9 | 10 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2009 | 16 | 16 | 18 | 14 | 14 | 18 | 15 | 21 | 24 | 27 |  |
| 2010 | 9 | 16 | 19 | 21 | 16 | 18 | 26 | 27 | 21 | 16 |  |
| 2011 | 20 | 24 | 27 | 19 | 23 | 17 | 25 | 23 | 23 | 35 |  |
| 2012 | 9 | 37 | 24 | 13 | 12 | 13 | 16 | 17 | 23 | 20 |  |
| 2013 | 17 | 17 | 15 | 23 | 20 | 21 | 16 | 17 | 31 | 38 |  |
| 2015 | 19 | 17 | 18 | 27 | 29 | 22 | 30 | 19 | 19 | 23 |  |
| 2016 | 20 | 13 | 13 | 10 | 9 | 13 | 16 | 24 | 20 | 20 |  |
| 2017 | 30 | 20 | 17 | 15 | 15 | 19 | 16 | 16 | 16 | 16 | 16 |

Table 2.2.9. Proportion Norwegian coastal cod by age among all aged cod in the Norwegian coastal survey north of $67^{\circ} \mathrm{N}$. The split between coastal cod and Northeast Arctic cod is uncertain for age 1.
$\left.\begin{array}{lllllllllll}\hline \text { Year } & \text { Age } & \mathbf{2} & \mathbf{3} & \mathbf{4} & \mathbf{5} & \mathbf{6} & \mathbf{7} & \mathbf{8} & \mathbf{9} & \mathbf{l} \\ \text { Total } \\ \text { number } \\ \text { of aged } \\ \text { cod oto- } \\ \text { liths }\end{array}\right\}$

| Year | Age 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | Total number of aged cod otoliths |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 0.73 | 0.81 | 0.76 | 0.82 | 0.73 | 0.61 | 0.69 | 0.43 | 0.83 | 0.50 | 1021 |
| 2008 | 0.99 | 0.99 | 0.99 | 0.83 | 0.89 | 0.84 | 0.78 | 0.67 | 0.94 | 0.75 | 1448 |
| 2009 | 0.94 | 0.94 | 0.83 | 0.69 | 0.55 | 0.58 | 0.75 | 0.76 | 0.73 | 0.72 | 1944 |
| 2010 | 0.94 | 0.94 | 0.89 | 0.75 | 0.66 | 0.49 | 0.60 | 0.86 | 0.90 | 0.97 | 2093 |
| 2011 | 0.90 | 0.93 | 0.91 | 0.89 | 0.77 | 0.66 | 0.52 | 0.73 | 0.80 | 0.83 | 1577 |
| 2012 | 0.94 | 0.89 | 0.90 | 0.82 | 0.83 | 0.73 | 0.71 | 0.61 | 0.88 | 0.84 | 1831 |
| 2013 | 0.93 | 0.94 | 0.88 | 0.77 | 0.79 | 0.83 | 0.74 | 0.79 | 0.73 | 1.00 | 1920 |
| 2014 | 0.99 | 0.99 | 0.99 | 0.96 | 0.93 | 0.90 | 0.93 | 0.87 | 0.87 | 0.88 | 2361 |
| 2015 | 0.89 | 0.93 | 0.89 | 0.86 | 0.75 | 0.73 | 0.65 | 0.73 | 0.82 | 0.96 | 1859 |
| 2016 | 0.99 | 0.98 | 0.99 | 0.90 | 0.84 | 0.69 | 0.75 | 0.80 | 0.71 | 0.83 | 2041 |
| 2017 | 1.00 | 0.98 | 0.95 | 0.93 | 0.86 | 0.74 | 0.78 | 0.68 | 0.84 | 1.00 | 1732 |
| 2018 | 0.99 | 0.97 | 0.91 | 0.86 | 0.88 | 0.82 | 0.72 | 0.68 | 0.87 | 0.90 | 2395 |
| 2019 | 0.95 | 0.99 | 0.97 | 0.88 | 0.84 | 0.83 | 0.84 | 0.76 | 0.82 | 0.91 | 2107 |
| 2020 | 1.00 | 0.84 | 0.85 | 0.81 | 0.71 | 0.70 | 0.75 | 0.83 | 0.78 | 0.64 | 2504 |

Table 2.2.10. Northern Norwegian coastal cod. Mean length (cm) at-age from Coastal survey data (A6335). Mean lengths of ages $>\mathbf{7}$ have higher uncertainty due to few samples. The split between coastal cod and Northeast Arctic cod is uncertain for age 1. For the plus group, mean length is the average mean length for ages 10+, weighted by abundance-at-age.

| Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1995 | 18.9 | 31.4 | 42.1 | 51.8 | 58.8 | 64.3 | 77.5 | 82.4 | 87.1 | 105.7 |
| 1996 | 16.7 | 28.3 | 41.3 | 51.9 | 58.1 | 65.2 | 74.8 | 86.7 | 99.6 | 115.0 |
| 1997 | 16.6 | 29.6 | 40.7 | 52.0 | 58.1 | 66.9 | 66.8 | 68.6 | 102.0 | 92.0 |
| 1998 | 17.8 | 30.3 | 44.0 | 52.0 | 60.3 | 67.8 | 74.9 | 82.2 | 83.8 | 107.8 |
| 1999 | 19.4 | 31.2 | 44.1 | 54.1 | 58.7 | 65.4 | 74.0 | 89.0 | 88.2 | 72.7 |
| 2000 | 20.0 | 32.5 | 44.0 | 54.0 | 61.4 | 64.5 | 73.8 | 81.9 | 80.3 | 90.3 |
| 2001 | 20.0 | 33.7 | 45.7 | 55.4 | 61.1 | 65.2 | 67.6 | 76.1 | 87.2 | 109.7 |
| 2002 | 21.6 | 32.6 | 45.0 | 54.5 | 62.0 | 68.8 | 72.4 | 70.5 | 66.7 | 91.8 |
| 2003 | 19.3 | 33.3 | 43.8 | 52.6 | 60.9 | 67.7 | 73.7 | 78.8 | 81.9 | 107.9 |


| Year | Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | 9 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 2.2.11. Northern Norwegian coastal cod. Mean weight (g) at-age from Coastal survey data (A6335). Mean weights of ages $>\mathbf{7}$ have higher uncertainty due to few samples. The split between coastal cod and Northeast Arctic cod is uncertain for age 1. For the plus group, mean weight is the average mean weight for ages 10+, weighted by abundance-at-age.

| Age |  | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1995 | 58 | 282 | 719 | 1395 | 2091 | 2767 | 4693 | 5905 | 7211 | 13022 |
| 1996 | 41 | 216 | 672 | 1349 | 1939 | 2779 | 4223 | 6638 | 11146 | 20000 |
| 1997 | 41 | 244 | 655 | 1393 | 1914 | 2921 | 2988 | 3768 | 9600 | 7779 |
| 1998 | 49 | 259 | 840 | 1406 | 2261 | 3173 | 4320 | 5275 | 5896 | 15476 |
| 1999 | 63 | 272 | 793 | 1508 | 1964 | 2759 | 4257 | 7262 | 6561 | 5934 |
| 2000 | 69 | 322 | 826 | 1561 | 2363 | 2811 | 4260 | 5977 | 6061 | 7553 |
| 2001 | 74 | 377 | 933 | 1660 | 2320 | 2998 | 3338 | 4478 | 7193 | 13677 |


| Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 2002 | 88 | 357 | 918 | 1595 | 2377 | 3468 | 4415 | 3868 | 3588 | 10135 |
| 2003 | 68 | 361 | 820 | 1427 | 2269 | 3127 | 4114 | 5493 | 6350 | 13767 |
| 2004 | 88 | 338 | 877 | 1646 | 2153 | 3197 | 3810 | 4656 | 4184 | 5457 |
| 2005 | 99 | 436 | 878 | 1727 | 2205 | 2542 | 3666 | 3520 | 5562 | 14216 |
| 2006 | 83 | 400 | 989 | 1649 | 2231 | 3502 | 3992 | 4445 | 8004 | 21921 |
| 2007 | 97 | 486 | 1066 | 1865 | 2579 | 3168 | 4520 | 6363 | 11111 | 13111 |
| 2008 | 97 | 427 | 1109 | 1971 | 3327 | 3393 | 4543 | 4921 | 4270 | 6451 |
| 2009 | 74 | 357 | 1032 | 1878 | 2695 | 3803 | 4599 | 5146 | 5349 | 5205 |
| 2010 | 63 | 502 | 1088 | 1872 | 2745 | 3586 | 4684 | 5096 | 6263 | 6698 |
| 2011 | 59 | 401 | 1165 | 2279 | 3109 | 3702 | 5163 | 5593 | 6174 | 5963 |
| 2012 | 73 | 355 | 1141 | 2026 | 2907 | 3690 | 4688 | 5549 | 6118 | 6504 |
| 2013 | 85 | 384 | 918 | 1817 | 3041 | 3438 | 3963 | 4926 | 5662 | 8265 |
| 2014 | 80 | 359 | 1122 | 1894 | 2929 | 3690 | 4646 | 5562 | 5550 | 8639 |
| 2015 | 73 | 406 | 1115 | 2145 | 2987 | 3774 | 4839 | 5299 | 5869 | 6708 |
| 2016 | 73 | 347 | 1101 | 1904 | 3327 | 3928 | 4689 | 5885 | 7273 | 8108 |
| 2017 | 83 | 504 | 1058 | 1969 | 2943 | 3997 | 4676 | 6985 | 6306 | 8472 |
| 2018 | 52 | 522 | 1109 | 2094 | 3206 | 3763 | 5391 | 5818 | 8438 | 6378 |
| 2019 | 62 | 372 | 1131 | 1984 | 2983 | 3815 | 5141 | 5908 | 6420 | 9215 |
| 2020 | 96 | 379 | 1010 | 1928 | 2972 | 3767 | 4995 | 5825 | 9305 | 7132 |

Table 2.2.12. Northern Norwegian coastal cod. Maturity-at-age as determined from maturity stages observed in the coastal survey (A6335). Maturity for age 10+ is the average proportion mature for ages 10 and above, weighted by abun-dance-at-age. The split between coastal cod and Northeast Arctic cod is uncertain for age 1.

|  | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1995 | 0.00 | 0.00 | 0.13 | 0.51 | 0.60 | 0.78 | 0.86 | 0.99 | 1.00 | 1.00 |
| 1996 | 0.00 | 0.02 | 0.14 | 0.38 | 0.74 | 0.84 | 0.92 | 1.00 | 1.00 | 1.00 |
| 1997 | 0.03 | 0.06 | 0.25 | 0.36 | 0.64 | 0.93 | 0.92 | 0.86 | 1.00 | 1.00 |
| 1998 | 0.01 | 0.03 | 0.13 | 0.24 | 0.56 | 0.70 | 0.98 | 0.93 | 0.88 | 1.00 |
| 1999 | 0.00 | 0.02 | 0.06 | 0.27 | 0.52 | 0.69 | 0.74 | 1.00 | 0.57 | 1.00 |
| 2000 | 0.00 | 0.00 | 0.06 | 0.20 | 0.51 | 0.68 | 0.80 | 0.92 | 1.00 | 1.00 |
| 2001 | 0.00 | 0.00 | 0.04 | 0.27 | 0.76 | 0.96 | 0.97 | 0.97 | 1.00 | 1.00 |
| 2002 | 0.00 | 0.01 | 0.11 | 0.30 | 0.78 | 0.89 | 0.98 | 0.94 | 1.00 | 1.00 |
| 2003 | 0.00 | 0.00 | 0.03 | 0.28 | 0.55 | 0.88 | 0.95 | 0.93 | 1.00 | 1.00 |
| 2004 | 0.00 | 0.01 | 0.11 | 0.30 | 0.78 | 0.92 | 0.94 | 1.00 | 1.00 | 1.00 |
| 2005 | 0.00 | 0.00 | 0.11 | 0.37 | 0.56 | 0.83 | 0.94 | 0.97 | 1.00 | 1.00 |
| 2006 | 0.00 | 0.01 | 0.19 | 0.53 | 0.72 | 0.93 | 0.90 | 0.96 | 1.00 | 1.00 |
| 2007 | 0.00 | 0.00 | 0.16 | 0.54 | 0.72 | 0.93 | 0.96 | 1.00 | 1.00 | 1.00 |
| 2008 | 0.00 | 0.02 | 0.10 | 0.30 | 0.73 | 0.88 | 0.97 | 1.00 | 1.00 | 1.00 |
| 2009 | 0.00 | 0.00 | 0.05 | 0.21 | 0.39 | 0.64 | 0.77 | 0.90 | 0.97 | 0.94 |
| 2010 | 0.00 | 0.00 | 0.03 | 0.27 | 0.57 | 0.78 | 0.92 | 0.99 | 0.98 | 1.00 |
| 2011 | 0.02 | 0.00 | 0.05 | 0.31 | 0.63 | 0.74 | 0.89 | 0.90 | 0.88 | 1.00 |
| 2012 | 0.00 | 0.01 | 0.04 | 0.28 | 0.57 | 0.86 | 0.89 | 1.00 | 0.96 | 1.00 |
| 2013 | 0.00 | 0.00 | 0.02 | 0.22 | 0.57 | 0.86 | 0.99 | 0.94 | 0.96 | 1.00 |
| 2014 | 0.00 | 0.00 | 0.03 | 0.15 | 0.56 | 0.78 | 0.90 | 0.98 | 1.00 | 1.00 |
| 2015 | 0.00 | 0.01 | 0.04 | 0.19 | 0.48 | 0.74 | 0.78 | 0.93 | 0.95 | 1.00 |
| 2016 | 0.00 | 0.00 | 0.06 | 0.28 | 0.61 | 0.85 | 0.91 | 0.98 | 1.00 | 1.00 |
| 2017 | 0.00 | 0.00 | 0.05 | 0.29 | 0.60 | 0.83 | 0.95 | 1.00 | 0.91 | 1.00 |
| 2018 | 0.00 | 0.00 | 0.07 | 0.24 | 0.60 | 0.79 | 0.94 | 1.00 | 1.00 | 1.00 |
| 2019 | 0.00 | 0.00 | 0.05 | 0.23 | 0.50 | 0.73 | 0.89 | 1.00 | 0.97 | 1.00 |
| 2020 | 0.00 | 0.02 | 0.07 | 0.33 | 0.61 | 0.88 | 0.97 | 0.98 | 1.00 | 1.00 |

Table 2.2.13. Northern Norwegian coastal cod. Tuning data used in the final SAM run.
Norwegian Coastal cod
101
A6335-acoustic-1995
19952002

| 1 | 1 | 0.75 | 0.85 |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 10 |  |  |  |  |  |  |  |  |
| 1 | 8.774 | 4.974 | 6.382 | 6.440 | 4.373 | 1.309 | 0.532 | 0.319 | 0.132 |
| 1 | 9.025 | 8.592 | 4.576 | 5.306 | 2.723 | 1.022 | 0.213 | 0.032 | 0.024 |
| 1 | 15.358 | 16.930 | 7.710 | 4.484 | 2.316 | 0.716 | 0.328 | 0.059 | 0.033 |
| 1 | 6.757 | 8.524 | 8.261 | 3.717 | 1.530 | 0.700 | 0.102 | 0.122 | 0.045 |
| 1 | 3.486 | 3.387 | 2.788 | 2.498 | 0.751 | 0.172 | 0.030 | 0.022 | 0.020 |
| 1 | 7.439 | 5.831 | 3.939 | 3.853 | 2.825 | 0.622 | 0.258 | 0.071 | 0.032 |
| 1 | 4.551 | 4.246 | 3.776 | 2.184 | 1.499 | 0.974 | 0.149 | 0.029 | 0.093 |
| 1 | 2.071 | 2.532 | 2.926 | 2.075 | 0.970 | 0.596 | 0.293 | 0.106 | 0.124 |

A6335-acoustic-2003

| 2003 | 2020 |  |
| :--- | :--- | :--- |
| 1 | 1 | 0.75 |
| 2 | 10 |  |
| 1 | 2.168 | 3.026 |
| 1 | 2.643 | 2.819 |
| 1 | 1.201 | 2.228 |
| 1 | 1.836 | 2.587 |
| 1 | 3.037 | 2.778 |
| 1 | 1.739 | 1.684 |
| 1 | 1.502 | 2.084 |
| 1 | 2.503 | 2.853 |
| 1 | 2.542 | 1.869 |
| 1 | 2.170 | 3.546 |
| 1 | 3.084 | 1.597 |
| 1 | 4.171 | 3.066 |
| 1 | 3.122 | 2.465 |
| 1 | 3.341 | 3.667 |
| 1 | 3.289 | 3.202 |
| 1 | 2.847 | 1.837 |
| 1 | 2.992 | 3.724 |
| 1 | 1.619 | 3.365 |


| A6335-trawl-2003 |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2003 | 2020 |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.75 | 0.85 |  |  |  |  |  |  |
| 2 | 10 |  |  |  |  |  |  |  |  |
| 1 | 3.268 | 3.763 | 4.521 | 2.700 | 2.319 | 0.863 | 0.489 | 0.220 | 0.069 |
| 1 | 2.201 | 2.396 | 2.602 | 1.463 | 0.722 | 0.359 | 0.181 | 0.046 | 0.063 |
| 1 | 1.042 | 1.988 | 1.478 | 1.268 | 0.746 | 0.157 | 0.107 | 0.068 | 0.054 |
| 1 | 2.156 | 2.623 | 2.946 | 1.554 | 1.026 | 0.941 | 0.171 | 0.107 | 0.023 |
| 1 | 0.911 | 0.853 | 1.071 | 0.789 | 0.465 | 0.394 | 0.114 | 0.075 | 0.029 |
| 1 | 1.822 | 2.795 | 1.883 | 1.419 | 1.145 | 0.580 | 0.348 | 0.161 | 0.094 |
| 1 | 2.251 | 3.570 | 3.716 | 1.584 | 0.868 | 0.712 | 0.466 | 0.204 | 0.160 |
| 1 | 2.353 | 3.268 | 3.385 | 2.397 | 0.784 | 0.383 | 0.733 | 0.317 | 0.328 |
| 1 | 3.471 | 2.498 | 2.866 | 2.095 | 1.445 | 0.292 | 0.315 | 0.213 | 0.310 |
| 1 | 3.218 | 4.485 | 2.784 | 1.537 | 1.042 | 0.930 | 0.411 | 0.200 | 0.346 |
| 1 | 4.101 | 1.706 | 2.666 | 1.887 | 1.575 | 0.890 | 0.578 | 0.297 | 0.419 |
| 1 | 5.448 | 4.026 | 3.034 | 3.521 | 2.016 | 1.388 | 0.465 | 0.364 | 0.337 |
| 1 | 4.733 | 4.154 | 3.727 | 2.068 | 1.818 | 0.902 | 0.506 | 0.397 | 0.222 |
| 1 | 4.433 | 4.522 | 2.610 | 1.995 | 0.746 | 0.735 | 0.413 | 0.203 | 0.210 |
| 1 | 2.891 | 2.407 | 1.563 | 1.151 | 0.715 | 0.308 | 0.200 | 0.147 | 0.157 |
| 1 | 3.197 | 1.916 | 1.879 | 1.049 | 0.748 | 0.323 | 0.183 | 0.128 | 0.168 |
| 1 | 2.114 | 2.470 | 1.508 | 1.460 | 0.839 | 0.490 | 0.148 | 0.129 | 0.211 |
| 1 | 1.670 | 2.599 | 2.416 | 1.188 | 0.611 | 0.291 | 0.177 | 0.049 | 0.072 |

Table 2.2.14. Northern Norwegian coastal cod. Stock mean weight-at-age (kg) was used in the assessment model. Mean weights at age in the catch are used in place of stock weights for ages 8-10+. Mean weights in 1994, when the survey had not yet started, are means of stock weights in the years 1995-1997 for ages 2-7 and set to weight in catch for ages 8-10+.

|  | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1994 | 0.247 | 0.682 | 1.379 | 1.981 | 2.822 | 3.968 | 5.245 | 6.487 | 8.825 |
| 1995 | 0.282 | 0.719 | 1.395 | 2.091 | 2.767 | 4.693 | 5.228 | 6.121 | 9.469 |
| 1996 | 0.216 | 0.672 | 1.349 | 1.939 | 2.779 | 4.223 | 4.544 | 5.462 | 7.814 |
| 1997 | 0.244 | 0.655 | 1.393 | 1.914 | 2.921 | 2.988 | 4.738 | 5.616 | 7.768 |
| 1998 | 0.259 | 0.840 | 1.406 | 2.261 | 3.173 | 4.320 | 4.786 | 5.389 | 9.584 |
| 1999 | 0.272 | 0.793 | 1.508 | 1.964 | 2.759 | 4.257 | 4.923 | 5.415 | 8.339 |
| 2000 | 0.322 | 0.826 | 1.561 | 2.363 | 2.811 | 4.260 | 5.553 | 5.834 | 9.781 |
| 2001 | 0.377 | 0.933 | 1.660 | 2.320 | 2.998 | 3.338 | 4.498 | 4.794 | 7.711 |
| 2002 | 0.357 | 0.918 | 1.595 | 2.377 | 3.468 | 4.415 | 5.268 | 6.236 | 9.943 |
| 2003 | 0.361 | 0.820 | 1.427 | 2.269 | 3.127 | 4.114 | 5.417 | 5.713 | 9.07 |
| 2004 | 0.338 | 0.877 | 1.646 | 2.153 | 3.197 | 3.810 | 5.367 | 5.93 | 7.991 |
| 2005 | 0.436 | 0.878 | 1.727 | 2.205 | 2.542 | 3.666 | 5.233 | 5.981 | 8.32 |
| 2006 | 0.400 | 0.989 | 1.649 | 2.231 | 3.502 | 3.992 | 5.806 | 6.638 | 9.71 |
| 2007 | 0.486 | 1.066 | 1.865 | 2.579 | 3.168 | 4.520 | 5.781 | 6.871 | 9.771 |
| 2008 | 0.427 | 1.109 | 1.971 | 3.327 | 3.393 | 4.543 | 5.844 | 6.279 | 9.239 |
| 2009 | 0.357 | 1.032 | 1.878 | 2.695 | 3.803 | 4.599 | 6.178 | 6.516 | 9.248 |
| 2010 | 0.502 | 1.088 | 1.872 | 2.745 | 3.586 | 4.684 | 5.437 | 6.185 | 7.599 |
| 2011 | 0.401 | 1.165 | 2.279 | 3.109 | 3.702 | 5.163 | 5.941 | 6.422 | 8.346 |
| 2012 | 0.355 | 1.141 | 2.026 | 2.907 | 3.690 | 4.688 | 6.448 | 6.914 | 9.446 |
| 2013 | 0.384 | 0.918 | 1.817 | 3.041 | 3.438 | 3.963 | 5.892 | 6.800 | 10.104 |
| 2014 | 0.359 | 1.122 | 1.894 | 2.929 | 3.690 | 4.646 | 5.791 | 6.461 | 9.643 |
| 2015 | 0.406 | 1.115 | 2.145 | 2.987 | 3.774 | 4.839 | 5.601 | 6.482 | 9.044 |
| 2016 | 0.347 | 1.101 | 1.904 | 3.327 | 3.928 | 4.689 | 5.893 | 6.850 | 8.928 |
| 2017 | 0.504 | 1.058 | 1.969 | 2.943 | 3.997 | 4.676 | 5.977 | 6.933 | 9.356 |
| 2018 | 0.522 | 1.109 | 2.094 | 3.206 | 3.763 | 5.391 | 5.711 | 6.581 | 9.333 |
| 2019 | 0.372 | 1.131 | 1.984 | 2.983 | 3.815 | 5.141 | 5.748 | 6.562 | 8.561 |


|  | Age |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| 2020 | 0.379 | 1.010 | 1.928 | 2.972 | 3.767 | 4.995 | 5.655 | 6.387 | 9.024 |

Table 2.2.15. Northern Norwegian coastal cod. Natural mortality at age is used in the assessment model. Estimated from mean weights at age (Table 2.2.14) by the Lorenzen (1996) method.

|  | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1994 | 0.687 | 0.504 | 0.407 | 0.364 | 0.327 | 0.295 | 0.271 | 0.254 | 0.231 |
| 1995 | 0.661 | 0.496 | 0.405 | 0.358 | 0.329 | 0.280 | 0.271 | 0.258 | 0.226 |
| 1996 | 0.716 | 0.507 | 0.410 | 0.367 | 0.329 | 0.289 | 0.283 | 0.267 | 0.240 |
| 1997 | 0.690 | 0.511 | 0.406 | 0.368 | 0.324 | 0.321 | 0.279 | 0.265 | 0.240 |
| 1998 | 0.677 | 0.473 | 0.404 | 0.350 | 0.316 | 0.287 | 0.278 | 0.268 | 0.225 |
| 1999 | 0.668 | 0.482 | 0.396 | 0.365 | 0.329 | 0.288 | 0.276 | 0.268 | 0.235 |
| 2000 | 0.634 | 0.476 | 0.392 | 0.345 | 0.327 | 0.288 | 0.266 | 0.262 | 0.224 |
| 2001 | 0.604 | 0.458 | 0.384 | 0.347 | 0.321 | 0.311 | 0.284 | 0.278 | 0.241 |
| 2002 | 0.615 | 0.461 | 0.389 | 0.345 | 0.307 | 0.285 | 0.270 | 0.257 | 0.223 |
| 2003 | 0.612 | 0.477 | 0.403 | 0.350 | 0.317 | 0.292 | 0.268 | 0.264 | 0.229 |
| 2004 | 0.625 | 0.467 | 0.386 | 0.355 | 0.315 | 0.298 | 0.269 | 0.261 | 0.238 |
| 2005 | 0.578 | 0.467 | 0.380 | 0.353 | 0.338 | 0.302 | 0.271 | 0.260 | 0.235 |
| 2006 | 0.594 | 0.450 | 0.385 | 0.351 | 0.306 | 0.294 | 0.262 | 0.252 | 0.224 |
| 2007 | 0.559 | 0.440 | 0.371 | 0.336 | 0.316 | 0.283 | 0.263 | 0.249 | 0.224 |
| 2008 | 0.582 | 0.435 | 0.365 | 0.311 | 0.309 | 0.283 | 0.262 | 0.256 | 0.228 |
| 2009 | 0.614 | 0.444 | 0.370 | 0.332 | 0.299 | 0.282 | 0.258 | 0.253 | 0.228 |
| 2010 | 0.554 | 0.437 | 0.371 | 0.330 | 0.304 | 0.280 | 0.268 | 0.257 | 0.242 |
| 2011 | 0.593 | 0.428 | 0.349 | 0.318 | 0.301 | 0.272 | 0.261 | 0.255 | 0.235 |
| 2012 | 0.615 | 0.431 | 0.362 | 0.324 | 0.301 | 0.280 | 0.254 | 0.249 | 0.226 |
| 2013 | 0.601 | 0.461 | 0.374 | 0.320 | 0.308 | 0.295 | 0.261 | 0.250 | 0.222 |
| 2014 | 0.613 | 0.433 | 0.369 | 0.323 | 0.301 | 0.281 | 0.263 | 0.254 | 0.225 |
| 2015 | 0.591 | 0.434 | 0.356 | 0.321 | 0.299 | 0.277 | 0.265 | 0.254 | 0.229 |
| 2016 | 0.620 | 0.436 | 0.369 | 0.311 | 0.296 | 0.280 | 0.261 | 0.250 | 0.230 |
| 2017 | 0.553 | 0.441 | 0.365 | 0.323 | 0.294 | 0.280 | 0.260 | 0.249 | 0.227 |


| Age |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | 2 | 3 | 4 | 0 | 0 | $10+$ |  |  |  |
| 2018 | 0.547 | 0.435 | 0.358 | 0.315 | 0.300 | 0.268 | 0.264 | 0.253 | 0.227 |
| 2019 | 0.607 | 0.432 | 0.364 | 0.322 | 0.298 | 0.272 | 0.263 | 0.253 | 0.233 |
| 2020 | 0.603 | 0.447 | 0.367 | 0.322 | 0.299 | 0.275 | 0.265 | 0.255 | 0.229 |

Table 2.2.16. SAM configuration.
Model used: SAM (State-space assessment model; https://www.stockassessment.org; Nielsen and Berg 2014).
Software used: Template Model Builder (TMB) and R.
Age range of assessment: $2-10$, where 10 is a plus group.
Start year of assessment: 1994
Last change of configuration: WKBarFar 2021
The assessment is available at www.stockassessment.org under the name NCCN67_AFWG2021_Corr
\# Configuration saved: Wed Jan 27 12:03:27 2021
\#
\# Where a matrix is specified rows corresponds to fleets and columns to ages.
\# Same number indicates same parameter used
\# Numbers (integers) starts from zero and must be consecutive
\#
\$minAge
\# The minimium age class in the assessment
2
\$maxAge
\# The maximum age class in the assessment
10
\$maxAgePlusGroup
\# Is last age group considered a plus group for each fleet (1 yes, or 0 no).
1111
\$keyLogFsta
\# Coupling of the fishing mortality states (nomally only first row is used).
012345556
-1 -1 -1 -1 -1 -1 -1 $-1 \begin{array}{ll}-1\end{array}$
-1 -1 -1 -1 -1 -1 -1 -1 -1
-1
\$corFlag
\# Correlation of fishing mortality across ages (0 independent, 1 compound symmetry, 2 AR(1), 3 separable AR(1)
\$keyLogFpar
\# Coupling of the survey catchability parameters (nomally first row is not used, as that is covered by fishing mortality).

- 1 - $1 \begin{array}{ccccccc} & -1 & -1 & -1 & -1 & -1 & -1\end{array}-1$

012334567
8910111111111112
131415161616161617
\$keyQpow

Table 2.2.16. SAM configuration continued.
\# Density dependent catchability power parameters (if any).

$$
\begin{aligned}
& \text {-1 }-1 \text {-1 }-1 \text {-1 }-1 \text {-1 }-1 \text {-1 } \\
& \begin{array}{cccccccc}
-1 & -1 & -1 & -1 & -1 & -1 & -1 & -1
\end{array} \text {-1 } \\
& \text {-1 }-1 \text {-1 }-1 \text {-1 }-1 \text {-1 }-1 \text {-1 }
\end{aligned}
$$

\$keyVarF
\# Coupling of process variance parameters for $\log (F)$-process (nomally only first row is used)
000000000
-1 -1 -1 -1 -1 -1 -1 -1 -1
-1

\$keyVarLogN
\# Coupling of process variance parameters for $\log (N)$-process
011111111
\$keyVarObs
\# Coupling of the variance parameters for the observations.
$\begin{array}{lllllllll}0 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$
222222222
$\begin{array}{lllllllll}3 & 3 & 3 & 3 & 3 & 3 & 3 & 3 & 3\end{array}$
$\begin{array}{lllllllll}4 & 4 & 4 & 4 & 4 & 4 & 4 & 4 & 4\end{array}$
\$obsCorStruct
\# Covariance structure for each fleet ("ID" independent, "AR" AR(1), or "US" for unstructured). | Possible values are: "ID" "AR" "US"
"ID" "ID" "AR" "AR"
\$keyCorObs
\# Coupling of correlation parameters can only be specified if the $A R(1)$ structure is chosen above.
\# NA's indicate where correlation parameters can be specified (-1 where they cannot).
\#2-3 3-4 4-5 5-6 6-7 7-8 8-9 9-10
NA NA NA NA NA NA NA NA
NA NA NA NA NA NA NA NA

```
01233445
67778999
```

\# Stock recruitment code (0 for plain random walk, 1 for Ricker, 2 for Beverton-Holt, and 3 piece-wise constant).
0

## \$noScaledYears

\# Number of years where catch scaling is applied.
0

Table 2.2.16. SAM configuration continued.

## \$keyScaledYears

\# A vector of the years where catch scaling is applied.
\$keyParScaledYA
\# A matrix specifying the couplings of scale parameters (nrow = no scaled years, ncols = no ages).
\$fbarRange
\# lowest and higest age included in Fbar
47
\$keyBiomassTreat
\# To be defined only if a biomass survey is used (0 SSB index, 1 catch index, 2 FSB index, 3 total catch, 4 total landings and 5 TSB index).
-1-1-1-1
\$obsLikelihoodFlag
\# Option for observational likelihood | Possible values are: "LN" "ALN"
"LN" "LN" "LN" "LN"
\$fixVarToWeight
\# If weight attribute is supplied for observations this option sets the treatment (0 relative weight, 1 fix variance to weight).
0
\$fracMixF
\# The fraction of $\mathrm{t}(3)$ distribution used in logF increment distribution
0
\$fracMixN
\# The fraction of $\mathrm{t}(3)$ distribution used in $\log \mathrm{N}$ increment distribution

0
\$fracMixObs
\# A vector with same length as number of fleets, where each element is the fraction of $t(3)$ distribution used in the distribution of that fleet

0000

## \$constRecBreaks

\# Vector of break years between which recruitment is at constant level. The break year is included in the left interval. (This option is only used in combination with stock-recruitment code 3)
\$predVarObsLink
\# Coupling of parameters used in a prediction-variance link for observations.

```
-1 -1 -1 -1 -1 -1 -1 -1 -1
-1 -1 -1 -1 -1 -1 -1 -1 -1
-1 -1 -1 -1 -1 -1 -1 -1 -1
-1 -1 -1 -1 -1 -1 -1 -1 -1
```

Table 2.2.17. SAM output. Estimated catchability at age for each fleet. In the SAM configuration, catchabilities are coupled (set equal) for ages 5-6 in the acoustic index part 1, and for ages 5-9 in the other two indices.

| Fleet | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Acoustic index pt. 1 | 0.103 | 0.163 | 0.228 | 0.308 | 0.308 | 0.233 | 0.128 | 0.097 | 0.126 |
| Acoustic index pt. 2 | 0.058 | 0.098 | 0.138 | 0.157 | 0.157 | 0.157 | 0.157 | 0.157 | 0.164 |
| Swept-area index | 0.060 | 0.100 | 0.140 | 0.153 | 0.153 | 0.153 | 0.153 | 0.153 | 0.173 |

Table 2.2.18. SAM output. Estimated recruitment (1000's), Spawning-stock biomass (SSB, t), average fishing mortalities for ages 4-7 (Fbar(4-7)), and Total-stock biomass (TSB, t).

| Year/Age | R (age 2) | Low | High | SSB | Low | High | Fbar (4-7) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 93167 | 64940 | 133663 | 121460 | 102525 | 143892 | 0.236 | 0.194 | 0.287 | 309739 | 270300 | 354933 |
| 1995 | 118218 | 86771 | 161062 | 102158 | 87017 | 119934 | 0.303 | 0.255 | 0.361 | 298854 | 264867 | 337203 |
| 1996 | 141681 | 103458 | 194025 | 80532 | 68850 | 94195 | 0.328 | 0.277 | 0.388 | 253383 | 224115 | 286474 |
| 1997 | 131307 | 96508 | 178652 | 65430 | 56184 | 76196 | 0.395 | 0.335 | 0.466 | 238844 | 208897 | 273083 |
| 1998 | 111445 | 82418 | 150695 | 56474 | 47967 | 66489 | 0.417 | 0.351 | 0.496 | 259744 | 225417 | 299299 |
| 1999 | 94384 | 69919 | 127409 | 46915 | 39424 | 55829 | 0.383 | 0.321 | 0.459 | 227431 | 197574 | 261800 |
| 2000 | 81853 | 60737 | 110309 | 53922 | 45426 | 64006 | 0.281 | 0.233 | 0.339 | 233031 | 202456 | 268225 |
| 2001 | 74792 | 55615 | 100582 | 69821 | 59298 | 82211 | 0.237 | 0.196 | 0.285 | 229830 | 199481 | 264797 |
| 2002 | 71973 | 54133 | 95692 | 83623 | 71350 | 98007 | 0.254 | 0.212 | 0.305 | 241969 | 211543 | 276771 |
| 2003 | 64546 | 50342 | 82760 | 70424 | 60137 | 82471 | 0.239 | 0.199 | 0.286 | 225252 | 197282 | 257187 |
| 2004 | 67260 | 53234 | 84980 | 74887 | 63786 | 87921 | 0.266 | 0.223 | 0.317 | 223234 | 194831 | 255779 |
| 2005 | 47688 | 36702 | 61962 | 66755 | 56611 | 78718 | 0.254 | 0.213 | 0.303 | 217546 | 189027 | 250368 |
| 2006 | 48613 | 37441 | 63119 | 83734 | 70408 | 99582 | 0.294 | 0.244 | 0.354 | 224622 | 195225 | 258446 |
| 2007 | 58323 | 45554 | 74671 | 88964 | 74190 | 106678 | 0.226 | 0.184 | 0.276 | 229070 | 197157 | 266148 |
| 2008 | 63129 | 49325 | 80798 | 86631 | 71479 | 104996 | 0.222 | 0.181 | 0.271 | 243754 | 208729 | 284657 |
| 2009 | 57062 | 44428 | 73289 | 67221 | 54983 | 82183 | 0.185 | 0.151 | 0.227 | 243099 | 207717 | 284509 |
| 2010 | 58091 | 45440 | 74264 | 81219 | 66377 | 99379 | 0.218 | 0.178 | 0.266 | 263740 | 225207 | 308866 |
| 2011 | 79004 | 62060 | 100573 | 92614 | 75146 | 114142 | 0.193 | 0.157 | 0.236 | 282675 | 240281 | 332550 |


| Year/Age | R (age 2) | Low | High | SSB | Low | High | Fbar (4-7) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 63136 | 49052 | 81263 | 98152 | 78499 | 122725 | 0.157 | 0.128 | 0.193 | 284079 | 239940 | 336337 |
| 2013 | 87542 | 68422 | 112006 | 104302 | 83762 | 129878 | 0.131 | 0.107 | 0.161 | 273810 | 231600 | 323712 |
| 2014 | 91802 | 72232 | 116674 | 110631 | 89811 | 136277 | 0.127 | 0.104 | 0.155 | 299085 | 254989 | 350806 |
| 2015 | 93654 | 73379 | 119530 | 100512 | 81617 | 123781 | 0.176 | 0.145 | 0.213 | 324205 | 278494 | 377419 |
| 2016 | 85893 | 67073 | 109994 | 108961 | 89125 | 133212 | 0.243 | 0.202 | 0.291 | 322536 | 277653 | 374675 |
| 2017 | 81129 | 62195 | 105829 | 92856 | 75042 | 114900 | 0.293 | 0.242 | 0.354 | 308476 | 262874 | 361989 |
| 2018 | 88742 | 66250 | 118869 | 87692 | 70095 | 109707 | 0.248 | 0.203 | 0.304 | 295124 | 245933 | 354154 |
| 2019 | 70293 | 50909 | 97057 | 77424 | 60225 | 99535 | 0.256 | 0.204 | 0.322 | 273093 | 221567 | 336601 |
| 2020 | 47259 | 31667 | 70530 | 80046 | 58135 | 110214 | 0.297 | 0.221 | 0.399 | 247612 | 191054 | 320911 |

Table 2.2.19. SAM output. Estimated fishing mortalities at age. F for ages $\mathbf{7 - 9}$ are coupled (set equal) in the SAM configuration.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 0.000 | 0.005 | 0.041 | 0.154 | 0.313 | 0.435 | 0.435 | 0.435 | 0.763 |
| 1995 | 0.000 | 0.008 | 0.056 | 0.183 | 0.382 | 0.593 | 0.593 | 0.593 | 1.003 |
| 1996 | 0.001 | 0.016 | 0.086 | 0.224 | 0.413 | 0.587 | 0.587 | 0.587 | 1.073 |
| 1997 | 0.001 | 0.021 | 0.112 | 0.27 | 0.509 | 0.689 | 0.689 | 0.689 | 1.296 |
| 1998 | 0.001 | 0.033 | 0.186 | 0.395 | 0.558 | 0.529 | 0.529 | 0.529 | 1.02 |
| 1999 | 0.001 | 0.026 | 0.158 | 0.363 | 0.499 | 0.513 | 0.513 | 0.513 | 0.982 |
| 2000 | 0.001 | 0.018 | 0.123 | 0.289 | 0.364 | 0.347 | 0.347 | 0.347 | 0.681 |
| 2001 | 0.001 | 0.014 | 0.094 | 0.225 | 0.313 | 0.314 | 0.314 | 0.314 | 0.723 |
| 2002 | 0.001 | 0.013 | 0.085 | 0.219 | 0.344 | 0.368 | 0.368 | 0.368 | 0.783 |
| 2003 | 0.001 | 0.012 | 0.066 | 0.185 | 0.311 | 0.393 | 0.393 | 0.393 | 0.759 |
| 2004 | 0.001 | 0.009 | 0.057 | 0.167 | 0.326 | 0.517 | 0.517 | 0.517 | 0.876 |
| 2005 | 0.001 | 0.009 | 0.059 | 0.169 | 0.295 | 0.496 | 0.496 | 0.496 | 0.949 |
| 2006 | 0.001 | 0.012 | 0.073 | 0.213 | 0.351 | 0.542 | 0.542 | 0.542 | 1.296 |
| 2007 | 0.001 | 0.017 | 0.078 | 0.2 | 0.271 | 0.351 | 0.351 | 0.351 | 0.845 |
| 2008 | 0.001 | 0.018 | 0.072 | 0.213 | 0.28 | 0.321 | 0.321 | 0.321 | 0.598 |
| 2009 | 0.001 | 0.017 | 0.051 | 0.163 | 0.254 | 0.27 | 0.27 | 0.27 | 0.457 |
| 2010 | 0.001 | 0.02 | 0.058 | 0.188 | 0.305 | 0.317 | 0.317 | 0.317 | 0.558 |
| 2011 | 0.002 | 0.024 | 0.067 | 0.154 | 0.232 | 0.317 | 0.317 | 0.317 | 0.49 |
| 2012 | 0.002 | 0.03 | 0.074 | 0.137 | 0.189 | 0.228 | 0.228 | 0.228 | 0.375 |
| 2013 | 0.002 | 0.029 | 0.07 | 0.113 | 0.15 | 0.189 | 0.189 | 0.189 | 0.311 |
| 2014 | 0.002 | 0.026 | 0.069 | 0.106 | 0.143 | 0.185 | 0.185 | 0.185 | 0.322 |
| 2015 | 0.003 | 0.034 | 0.091 | 0.137 | 0.202 | 0.269 | 0.269 | 0.269 | 0.465 |
| 2016 | 0.003 | 0.033 | 0.105 | 0.156 | 0.276 | 0.425 | 0.425 | 0.425 | 0.598 |
| 2017 | 0.003 | 0.04 | 0.126 | 0.196 | 0.317 | 0.512 | 0.512 | 0.512 | 0.638 |
| 2018 | 0.002 | 0.026 | 0.089 | 0.16 | 0.261 | 0.452 | 0.452 | 0.452 | 0.496 |
| 2019 | 0.002 | 0.023 | 0.089 | 0.161 | 0.275 | 0.448 | 0.448 | 0.448 | 0.467 |
| 2020 | 0.002 | 0.019 | 0.087 | 0.18 | 0.345 | 0.479 | 0.479 | 0.479 | 0.395 |

Table 2.2.20. SAM output. Estimated stock numbers at age (1000's).

| Year/Age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1994 | 93255 | 32066 | 35997 | 38592 | 18199 | 10235 | 4682 | 1160 | 1682 |
| 1995 | 117847 | 43217 | 21285 | 23442 | 21638 | 9340 | 4867 | 2455 | 1239 |
| 1996 | 140946 | 62223 | 25078 | 14316 | 13935 | 10251 | 3764 | 1861 | 1291 |


| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 130577 | 76812 | 33057 | 14449 | 8006 | 6968 | 4316 | 1535 | 1192 |
| 1998 | 111048 | 65408 | 46677 | 18559 | 7873 | 3358 | 2558 | 1496 | 835 |
| 1999 | 94243 | 53821 | 36124 | 24080 | 8125 | 3094 | 1542 | 1186 | 895 |
| 2000 | 81710 | 49201 | 31714 | 20474 | 11365 | 3591 | 1485 | 719 | 846 |
| 2001 | 74734 | 41855 | 31452 | 18267 | 10013 | 5635 | 1776 | 727 | 977 |
| 2002 | 72109 | 41589 | 25923 | 20052 | 10495 | 5119 | 2982 | 922 | 844 |
| 2003 | 66126 | 41601 | 28887 | 15395 | 12422 | 5415 | 2649 | 1549 | 736 |
| 2004 | 68658 | 37563 | 28319 | 17277 | 10227 | 5545 | 2664 | 1165 | 1044 |
| 2005 | 49178 | 43471 | 23381 | 18921 | 12501 | 4680 | 2573 | 1072 | 879 |
| 2006 | 50300 | 33063 | 24940 | 14636 | 11542 | 7402 | 2377 | 1267 | 679 |
| 2007 | 59807 | 28800 | 23273 | 15042 | 8997 | 6496 | 2676 | 1270 | 662 |
| 2008 | 63708 | 37949 | 19540 | 12887 | 9941 | 5312 | 3303 | 1297 | 980 |
| 2009 | 58336 | 39771 | 27528 | 12939 | 7403 | 5364 | 3147 | 2060 | 1072 |
| 2010 | 59296 | 37656 | 25472 | 19318 | 7767 | 4240 | 3415 | 1869 | 2063 |
| 2011 | 80455 | 32046 | 24763 | 15335 | 11639 | 4222 | 2578 | 1853 | 2191 |
| 2012 | 65191 | 50675 | 23464 | 14011 | 9968 | 6579 | 2607 | 1268 | 2478 |
| 2013 | 89485 | 27603 | 26457 | 15573 | 10630 | 6712 | 4150 | 1750 | 2129 |
| 2014 | 94428 | 37965 | 18091 | 19356 | 10407 | 7029 | 4131 | 2816 | 2261 |
| 2015 | 96804 | 43735 | 25607 | 12890 | 12932 | 6520 | 3976 | 2866 | 2879 |
| 2016 | 90624 | 49582 | 23692 | 18356 | 8179 | 7701 | 4240 | 2250 | 3128 |
| 2017 | 86788 | 48234 | 24428 | 16422 | 10856 | 5190 | 3687 | 2207 | 2376 |
| 2018 | 97956 | 37373 | 28342 | 14221 | 10350 | 5537 | 2711 | 1695 | 1945 |
| 2019 | 83054 | 53667 | 21894 | 17808 | 10098 | 6042 | 2531 | 1470 | 1699 |
| 2020 | 54381 | 49334 | 34278 | 16180 | 9809 | 5644 | 3134 | 1117 | 1364 |

Table 2.2.21a. Northern Norwegian coastal cod. Assumptions for the interim year and in the forecast: Fbar, recruitment, SSB and catch.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| $\mathrm{F}_{\text {ages 4-7 (2021) }}$ | 0.275 | $\mathrm{~F}_{\text {sq }}=$ median fishing mortality in 2020. |
| SSB (2021) | 92885 | Short-term forecast fishing at status quo <br> ( $\mathrm{Fq}_{\text {sq }}$ ); Tonnes. |
| $\mathrm{R}_{\text {age 2 }}(2021,2022$, and 2023) | 88137 | Median resampled recruitment (2011- <br> 2020) as estimated by a stochastic <br> projection; Thousands. |
| Total catch (2021) | 47809 | Short-term forecast fishing at $\mathrm{F}_{\text {sq }} ;$ <br> Tonnes. |

Table 2.2.21b. Northern Norwegian coastal cod. Assumptions for the interim year and in the forecast: mean weights in catch and stock, maturity at age, and natural mortality at age ( 5 -year averages).

| Age | Weight in catch (kg) | Weight in stock (kg) | Proportion mature | Natural mortality |
| :--- | :--- | :--- | :--- | :--- |
| 2 | 1.331 | 0.425 | 0.006 | 0.586 |
| 3 | 1.862 | 1.082 | 0.059 | 0.438 |
| 4 | 2.515 | 1.976 | 0.273 | 0.365 |
| 5 | 3.251 | 3.086 | 0.582 | 0.318 |
| 6 | 4.052 | 3.854 | 0.815 | 0.297 |
| 7 | 4.890 | 4.978 | 0.933 | 0.275 |
| 8 | 5.797 | 5.797 | 0.991 | 0.263 |
| 9 | 6.663 | 9.663 | 0.976 | 0.252 |
| $10+$ | 9.040 |  | 1.000 |  |

Table 2.2.22. Northern Norwegian coastal cod. Catch scenarios.

| Basis | Total catch (2022) | Ftotal (2022) | SSB (2023)* | \% SSB change **\% Advice change *** |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| ICES advice basis |  |  |  |  |  |
| Precautionary approach | 7865 | 0.039 | 115782 | 25 | - |
| Other scenarios |  |  |  |  |  |
| $\mathrm{F}=0$ | 0 | 0 | 120404 | 30 | - |
| $\mathrm{F}=\mathrm{F}_{2020}$ | 48497 | 0.275 | 92748 | -0.15 | - |
| $\mathrm{F}=0.1^{\wedge}$ | 19435 | 0.10 | 109084 | 17 |  |

* For this stock, SSB is calculated at the time of survey (October) as maturity ogives and stock weights are from the survey. Thus, SSB is influenced by fisheries between 1 January and 1 October. The actual spawning time is MarchJune.
** SSB in October 2022 relative to SSB in October 2021.
*** Advice value for 2022 relative to advice value for 2021. Not presented this year as it is the first advice for this stock.
${ }^{\wedge}$ Corresponding to the target $F$ in 2021 according to the previous management plan for the combined northern and southern coastal cod.


Figure 2.2.1. Northern Norwegian coastal cod. Standard figures. SAM estimates of a) SSB, b) Fbar(4-7), c) recruitment (age 2,), and d) catch input data.


Figure 2.2.2. Acoustic abundance index by age (colours) from the Coastal survey in October-November (survey code A6335).


Figure 2.2.3. Swept area abundance index by age (colours) from the coastal survey in October-November (survey code A6335).


2.2.4. Survey mortality $(Z)$ at age (colours) in the acoustic index (top) and swept area index (bottom). $Z$ was estimated as $-\log \left(A_{a+1, y+a} / A_{a, y}\right)$, where $A_{a, y}$ is abundance of age $a$ in year $y$.


Figure 2.2.5. Mean weight-at-age in the coastal survey. Few individuals of ages 10+ were sampled in the beginning of the time series, leading to extremely large variation in mean weights.


Figure 2.2.6. Proportions mature-at-age as observed in the Coastal survey.

2.2.7. Natural mortality-at-age estimated from stock weights-at-age by the Lorenzen (1996) method.


Figure 2.2.8. Northern Norwegian coastal cod. 5-year retrospective peel: a) SSB, b) Fbar, c) recruitment, and d) catch. The Mohn's rho value (average retrospective bias) is indicated in the upper right corner of each panel.


Figure 2.2.9. Residuals for the $\log (N)$ (top) and $\log (F)(b o t t o m)$ process from the final SAM run.


Figure 2.2.10. One-step-ahead residuals by fleet from the final SAM run. Blue circles indicate positive residuals and red circles indicate negative residuals. Top left: catch, top right: acoustic index pt. 1, bottom left: acoustic index pt. 2, bottom right: swept-area index.


Figure 2.2.11. Stock-recruitment relationship from SAM. Estimated recruitment-at-age $\mathbf{2}$ ( $\mathbf{1 0 0 0}$ 's) is plotted against estimated SSB ( t ) in the year of spawning (two years previously). The year labels in the figure indicate year of recruitment.


Figure 2.2.12. Comparative stock-recruitment relationship: estimated abundance-at-age 3 (1000’s) plotted against estimated SSB ( $\mathbf{t}$ ) in the year of spawning (three years previously). Recruitment in $\mathbf{2 0 2 0}$ is marked with a red triangle.


Figure 2.2.13. Short-term prediction. Predicted SSB (top panels), Fbar (middle panels) and recruitment (bottom panels) at status quo fishing (top left), status quo then zero fishing (top right), fishing at the level that will put the stock above $B_{\text {lim }}$ at the end of the advice year (bottom left), and $F=0.1$, current $F$ target in the old management plan for all coastal cod north of $62^{\circ} \mathbf{N}$ (northern and southern Norwegian coastal cod). In the forecast, recruitment is the same for all scenarios (resampled from the period 2003-2020).

### 2.3 Coastal cod south between 62-67º N (Southern Norwegian coastal cod)

### 2.3.1 Stock status summary

An assessment based on the decisions of the 2021 WKBARFAR benchmark (ICES 2021a) is presented for this stock.

The catches have decreased since 2010-2012, to a large extent explained by a decreased commercial fishing effort until 2017 but have continued to decrease even after 2017 when the effort has been slightly increasing. The recreational fishery by tourists and Norwegian residents is assumed to catch similar amounts as the commercial fishery, and a prerequisite for more accurate future assessments is a better estimation of the recreational catches.

Until we have several years in the CPUE series and can use the recommended SPiCT or JABBA surplus production models, the assessment of coastal $\operatorname{cod} 62-67^{\circ} \mathrm{N}$ is trend-based (the "2 over 3" rule) using the Reference fleet CPUE (which is more controlled than a full fleet CPUE). LBSPR and other length-based indicators have been used as additional information to assess the need for a $20 \%$ precautionary buffer in the " 2 over 3 " rule. ICES lacks for time being a framework for using LBSPR directly as a basis for quota advice.

Between 2007-2019, the mean "Spawning potential ratio", i.e. the ratio between the recruitment potential of the current stock and the theoretical recruitment potential without fishing, fluctuated between 20 and $30 \%$, with an overall downward trend. This places the stock below the target values ( $30-40 \%$ ) and - at the end of the series - even below $20 \%$, generally accepted as a limit reference point in the absence of further information on the stock dynamics. The decrease in the spawning potential ratio is concomitant with a decline of both mean length and mean length of the largest $5 \%$ of the caught fish. These all together depict a somewhat depleted and worsening stock status.

The ratio between the two last year's CPUE (2019-2020) and the three previous years (2016-2018) gives a factor of 1.17 . Including a precautionary $20 \%$ results in a final factor of 0.94 , or a recommended $6 \%$ decrease in catch advice compared to the three last years' catches.

No previous advice has been issued for this stock. The 2021 advice for the previous Norwegian coastal cod stock (comprising the two new stocks) was to follow the Norwegian management plan, which implied reducing fishing mortality to 0.05 .

The new formal name of the stock is "Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ (Norwegian coastal cod South)" and its stock code "cod.27.2.coastS".

### 2.3.2 Fisheries (Table 2.3.2-Table 2.3.4)

Coastal cod is fished throughout the year but the main (about 70\%) commercial fishery for coastal cod in the area between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ takes place during February-April. The main fishing areas are along the coast of Helgeland including Træna and Lovund, Vikna, Halten bank, and further along the coast of Trøndelag and Møre and Romsdal counties. Except for the Borgundfjord at Møre, the quantities fished inside fjords are quite low.
In the 1990ies the average percentage share between gear types in the estimated coastal cod commercial landings was around $65 \%$ for gillnet, $26 \%$ for longline/handline, $8 \%$ for Danish seine, and $1 \%$ for bottom trawl. In 2020 this share was $67 \%$ for gillnet, $30 \%$ for longline/handline/Danish seine, and $3 \%$ for bottom trawl (Table 2.3.4).

Recreational and tourist fisheries take an important fraction of the total catches in some local areas, especially near the coastal cities, and in some fjords where commercial fishing activity is low. However, there are a few reports trying to assess the amount in certain years (see section 2.1). The current split of the recreational catches between the area north of $67^{\circ} \mathrm{N}$ and between $62-$ $67^{\circ} \mathrm{N}$ in 2019-2020 is done based on the tourist fishing businesses' reporting to the Norwegian Directorate of Fisheries by county. Since the $67^{\circ} \mathrm{N}$ latitude goes through the Nordland county, the splitting north and south of $67^{\circ} \mathrm{N}$ for this county is done proportional to the number of tourist fishing businesses north and south of this latitude. The same area proportion ( $37.8 \%$ south and $62.2 \%$ north) of the recreational fishery is used for the whole time-series back to 1994, and this is a very rough assumption that should be further investigated and better documented. In recent years the recreational cod catches between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ are estimated to about $55 \%$ of total cod catches in this region (Tables 2.1.1 and 2.3.3).

Discarding is known to take place. There have previously been conducted two investigations trying to estimate the level of discarding and misreporting from coastal fishing vessels in two periods (2000 and 2002-2003, WD 14 at 2002 WG). The amount of discards was calculated, and the report from the 2000 -investigation concluded there was both discard and misreporting by species in 2000, in the gillnet fishery approximately $8-10 \%$ relative to reported catch. $1 / 3$ of this was probably coastal cod. The last report concluded that misreporting in the Norwegian coastal gillnet fisheries have been reduced significantly since 2000.

According to a recent report by Berg and Nedreaas (2021) up to 5\% was discarded in the commercial gillnet fishery between $62-67^{\circ} \mathrm{N}$ during $2012-2018$, and about $7 \%$ in the rod and line sector of the recreational fishery. The latter estimate is based on reporting to the Directorate of Fisheries in 2019 showing that about $35 \%$ of the reported rod and line catch was released with an assumed mortality of $20 \%$ of the released cod (see section 2.1). Discarding is not included in the commercial catch in this report but discarding in the rod and line (from boat) sector of the recreational fishery is included in the recreational catch estimate.

### 2.3.2.1 Estimated catches and Catch-at-age (Table 2.3.2-Table 2.3.4, and Figure

### 2.1.1 and Figure 2.3.1-Figure 2.3.2)

The current coastal cod assessments include all coastal cod caught within the coastal statistical areas 600, 601, 700 and 701 which extend beyond the 12 nautical mile zone (see Figure 2.1.1). Estimated commercial and recreational catches of Coastal cod and Northeast Arctic cod in these statistical areas between $62-67^{\circ} \mathrm{N}$ are shown in Table 2.1 and in Figures 2.3.1-Figure 2.3.2.

The estimated commercial catch-at-age (2-10+) for the period 1994-2020 is given in Table 2.3.2. Table 2.3.3 shows the total catch numbers-at-age when recreational and tourist fishing is included. The commercial catch in 2020 by gear and Norwegian statistical fishing areas is presented in Table 2.3.4.

### 2.3.2.2 Catch weights-at-age (Table 2.3.5)

Weight-at-age in catches is derived from the commercial sampling and is shown in Table 2.3.5. The same weight-at-age is assumed for the recreational and tourist catches.

### 2.3.2.3 Catches in 2021

No catch prediction for 2021 have been made, but it is reasonable to assume the same catch level as in 2020, i.e. a somewhat reduced recreational fishery due to the Covid19 pandemic and travel restrictions for foreign tourists.

### 2.3.3 Reference fleet

The Norwegian Reference Fleet is a group of active fishing vessels paid and tasked with providing information about catches (self-sampling) and general fishing activity to the Institute of Marine Research. The fleet consists of both high seas and coastal vessels that cover most of the Norwegian waters. The Highseas Reference Fleet began in 2000 and was expanded to include coastal vessels in 2005 (Clegg and Williams, 2020). The Coastal reference fleet has reported catch-pergillnet soaking time (CPUE) from their daily catch operations (WD 07).

These fleets catch cod from both coastal and NEA populations, which can be discriminated based on their otolith shape. Size distribution of individuals is sampled from a subset of fishing events and, within the size samples, individuals are sampled for otolith in a presumably random way.

To determine the origin of the cod, we use all data from north of $62^{\circ} \mathrm{N}$ (i.e. ICES Subarea 2.a.2; Norwegian statistical areas $3,4,5,0,6,7$ ) with information on otolith type. The probability of a fish caught to be coastal cod (as opposed to NEA cod) is modelled using a Binomial GLM. The covariates area (Norwegian statistical area), year, quarter and gear, all coded as factors, were examined and a model selection was performed based on an information theory approach. The modelled proportions of coastal cod per area and quarter, from 2007 to 2020, are presented in the Stock Annex. Further use for the elaboration of the CPUE index specifically focuses on areas 6 and 7 (between $62-67^{\circ} \mathrm{N}$ ) and quarters 3 and 4 because it is believed that this is the best data to inform about coastal cod status in this area.

### 2.3.4 CPUE standardization of reference fleet data (Table 2.3.6 and Figure 2.3.3-Figure 2.3.7.

Raw CPUE data are seldom proportional to population abundance as many factors (e.g. changes in fish distribution, catch efficiency, effort, etc) potentially affect its value. Therefore, CPUE standardization is an important step that attempts to derive an index that tracks relative population dynamics.

There are two cod stocks (two ecotypes) that are mixed in the Norwegian waters: the coastal cod (NCC) and the Northeast Arctic cod (NEAC). In this working document, our interest lies in deriving the abundance index of coastal cod, therefore, a few steps need to be taken to derive the corresponding coastal cod abundance index:

1. Fit a model to determine whether an individual fish is categorized as coastal or NEAC. This step allows determining the probability of catching coastal cod vs. NEAC during the time frame of interest.
2. Perform a CPUE standardization using the data from the reference fleet (on total cod catch; the division to ecotypes happens in the next step).
3. Use the output from the above two steps and create an index of abundance for coastal cod.

Below, we defined some important terms we used for the CPUE standardization.
Standardized effort (gillnet day) = gear count x soaking time (hours) / 24 hours
CPUE (per gillnet day) = catch weight/standardized effort

## Step 1: Coastal cod vs. NEAC?

In order to determine the origin of cod, we used all data from above $62^{\circ} \mathrm{N}$ (i.e. areas $3,4,5,0,6$, 7) with information on otolith type. The latter is the source of identification that helps separate
coastal vs. NEAC. Otolith types 1 and 2 were categorized as "coastal" and type 3, 4, 5, as NEAC. A total of 27897 samples were used for the analysis between 2007-2020.

From the above samples, we removed any covariates that had less than three observations to ensure estimability (the covariate in question was mostly the gear type; the final sample size was $\mathrm{N}=27892$ ). We then fitted a binomial model with logit link using four different explanatory variables: year, area, quarter, and gear, using the following formula:

$$
\begin{aligned}
& \text { Glm1 <- glm(is_coastal ~ factor(area)*factor(startyear) }+ \text { factor(quarter) }+ \text { factor(gear), fam- } \\
& \text { ily=binomial, data=Data_proportion) }
\end{aligned}
$$

Using the above model (Figure 2.3.3), we then predicted the proportion of coastal cod that would be expected in areas 6 and 7, during quarters 3 and 4, between 2007-2020 (see Figure 2.3.4).

## Step 2: CPUE standardization

Many different R packages (e.g. mgcv::gam, glmmTMB::glmmTMB, sdmTMB::sdmTMB, and own model in TMB to allow implementing a mixture model), as well as many different combinations of likelihood functions (e.g. normal, lognormal, gamma, negative binomial, student t , tweedie), zero inflation, and parameter, were tested to find a model which showed an acceptable residual pattern. However, model exploration was not conclusive when using the entire CPUE data from the area north of $62^{\circ} \mathrm{N}(\mathrm{N}=11805$, with only 59 zeros $)$. All the models struggled to fit the extremely skewed CPUE data (many extremely small values below 1 and large values above 1000, while the bulk of the values are in the scale of dozens).

The final model for the CPUE standardization was fitted on all cod data (no distinction between coastal and NEAC yet) but limited to areas 6 and 7 and quarters 3 and 4, between 2007-2020. Further data filtering was performed to remove erroneous data points (e.g. gearcount $=1$ ) and any gear code with less than 3 observations or only used in one year. This reduced the final data set to $\mathrm{N}=686$ (with only 3 zeros):

$$
\begin{align*}
& \text { glmmTMB_pos <- glmmTMB(log(cpue_all) } \sim \text { factor(startyear })+ \text { factor(area) }+ \text { factor(gear) }+ \\
& \text { factor(quarter })+(1 \text { larea_year })+(1 \mid \text { quarter_year }), \text { family }=\text { gaussian, data=subset(nord_use, } \\
& \text { cpue_all>0) }) \tag{eq2}
\end{align*}
$$

The expression (1|area_year) indicates that the area and year variable was concatenated into a single variable and considered as a random effect acting on the intercept. In essence, this treatment models the interaction effect between year and area on the intercept, but the approach only considers existing interaction (as opposed to all possible combinations of year and area which would be un-estimable) - which is an advantage in a data-limited situation such as ours.

## Joining steps 1 and 2 to create a standardized coastal cod CPUE

The final cod CPUE model showed a reasonable residual behaviour (Figure 2.3.5) and therefore, we proceeded with the derivation of the standardized coastal cod CPUE index for areas 6 and 7 and quarters 3 and 4 .

The standardized coastal cod index (CPUE_stdcoastal) was calculated as:
CPUE_stdcoastal = Pcoastal * CPUEcod

Where Pcoastal is the predicted proportion of coastal cod in the catch based on the output from step1, and CPUEcod is the predicted cod (of both ecotypes) CPUE based on step 2.

And the variance of (CPUE_stdcoastal) was calculated as:

$$
\begin{equation*}
V\left(C P U E_{-} s t d_{\text {coastal }}\right)=\left(\widehat{P_{\text {coastal }}}\right)^{2} V\left(C P U E_{\text {cod }}\right)+\left(\widehat{C P E_{c o d}}\right)^{2} V\left(P_{\text {coastal }}\right) \tag{eq4}
\end{equation*}
$$

Some combinations of area_year and quarter_year random interaction effect were not present in the datasets for the CPUE standardization model. However, glmmTMB can handle any missing new levels of random effect variables when making a prediction (it assumes it is equal to zero and inflates the prediction error by its associated random effect variance). For diagnostic plots, see WD 07.

The standardized CPUE index for coastal cod in areas 6 and 7, i.e. between $62-67^{\circ} \mathrm{N}$, during quarters 3 and 4, between 2007-2020, is shown in Figure 2.3.6. The composite standardized CPUE index for coastal cod in the entire area between $62-67^{\circ} \mathrm{N}$ during quarters 3 and 4 , is shown in Figure 2.3.7 and Table 2.3.6.

### 2.3.5 Stochastic LBSPR (Table 2.3.1)

Given the uncertainty in parameters and the demonstrated sensitivity of the model to input parameters (Hordyk et al., 2015b, 2015a), the AFWG has implemented a stochastic Length-based spawning potential ratio (LBSPR) approach similar on the principle to the one developed for anglerfish within the Arctic fisheries working group (see section 9). Differences with this former approach include variations in the parameterization of random inputs, and the inclusion, in the present model, of bootstrapped size distributions to account for uncertainty in the observation of length compositions.

Size distributions are estimated based on reference fleet data using, unlike for the CPUE index (see above), only catches sampled for size.

Most of the parameters estimated during WKBARFAR (ICES 2021) do not need to be re-evaluated on an annual basis and can be randomly generated using the mean and standard deviation from Table 2.3.1 below. Only in case of shift in the growth and/or condition of the fish should the growth parameters and/or the two natural mortality parameters (M and Mpow, sensitive to the conditions) be respectively re-estimated. Because they are more variable and have typically asymmetric distributions, it is recommended to regenerate sets of random maturity ogive each time with updated data.

Table 2.3.1. Parameters used to set up the stochastic LBSPR approach and their value (including uncertainty). Parameters in bold are the inputs of the LBSPR model. Other parameters not detailed here were left to their default values.

| Parameter | Mean value <br> (sd) | Description, comment |
| :--- | :--- | :--- |
| $M$ | $0.228(0.0012)$ | Natural mortality (year <br> mates based on resampled reference fleet commercial sampling data following Lo- <br> renzen (1996). |
| $M_{\text {pow }}$ | $0.939(0.0042)$ | aka exponent c, equ. 17 in Hordyk et al. (2016): parameterization of the size varying <br> mortality in LBSPR. Fitted from size varying M estimates, following Lorenzen (1996), <br> based on resampled reference fleet commercial sampling data. |
| $k$ | $0.248(0.0033)$ | growth coefficient from a von Bertalanffy growth function. |
| $M / k$ | $0.919(0.0078)$ | M/k at Los, derived from the above estimates. |
| $L_{\text {inf }}$ | $95.45(0.528) *$ | Asymptotic length $L_{\infty}(c m)$, as defined in a von Bertalanffy growth function. |


| Parameter | Mean value <br> (sd) | Description, comment |
| :--- | :--- | :--- |
| $\mathrm{t}_{0}$ | -0.0388 | Theoretical time (year) when length $=0$ in a von Bertalanffy growth function. Not a <br> LBSPR parameter per se, but used for the estimation of $k$ and Linf above parame- <br> ters. Estimate borrowed from the coastal cod North of $67^{\circ} \mathrm{N}$ (EP method). |
| CVLinf $^{\text {LM50 }}$ | $6.155(0.0006)$ | Coefficient of variation of asymptotic length. Encompass all inter-individual growth <br> variability of LBSPR. The values used are the CV of size at age, and its uncertainty, <br> estimated for the coastal cod North of $67^{\circ} \mathrm{N}$ (EP method). Estimated and randomly <br> generated on the log scale (mean $=-1.862 ;$ s.d. $=0.0039)$. |
| LM95 (1.688) + | Length (cm) at 50\% maturity. Estimated from resampled coastal survey data (2010- <br> 2019) using a binomial glm. |  |
| $79.92(3.924)+$ | Length (cm) at 95\% maturity. Estimated from resampled coastal survey data (2010- <br> 2019) using a binomial glm. |  |

*randomly generated preserving the correlation structure between $k$ and Linf using a multinormal distribution.
†pairs (LM50, LM95) estimated from a same bootstrapped dataset and year drawn together to preserve the correlation between the two parameters and avoid using a parameterization based on the distribution of $\Delta \mathrm{Lm}=\mathrm{LM} 95-\mathrm{LM} 50$.

## Growth parameters

In a von Bertalanffy growth model, the asymptotic length $(\mathrm{L} \infty)$ and the growth coefficient $(\mathrm{k})$ have strongly correlated estimates. This correlation should therefore be maintained when generating random parameters. This can be achieved using a multinormal distribution random generator with the means in Table 2.3.1 and the variance-covariance matrix in Stock Annex.

## Natural mortality

One of the most critical parameters for the performance of LBSPR is M/k. Here we had first-hand growth parameter estimates but no a priori information on M/k in coastal cod. Estimating M based on life history was therefore favoured and four methods tested: one giving a constant M (Then et al., 2015, 2018) and three size varying M estimates (Lorenzen, 1996; Gislason et al., 2010; Charnov et al., 2013). SPR estimates based on these four different M were shown to have different absolute values but fairly similar trends. Among the four options examined for the parameterization of natural mortality, the size varying M following Lorenzen (1996) was retained based on its consistency with cannibalism-driven mortality in the partially sympatric NEA cod. It also provides the SPR and $\mathrm{F} / \mathrm{M}$ estimates the closest to a $\mathrm{M}=0.2$ scenario, while there is consensus that it represents a more realistic alternative than the later.

The Lorenzen M estimate is based on individual weights but is here re-parameterized as length varying using individuals sampled for weight and length in the reference fleet data. It may therefore need to be re-estimated in case of sustained substantial shift in the condition of fish.

## Maturity ogive

Maturity is estimated for the whole autumn coastal survey data north of $62^{\circ} \mathrm{N}$, on account of scarcity of biological cod samples for the area between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ alone. For consistency with the choices made for the northern stock, resting individuals (stage 4) are included in the mature fraction. The maturity parameters (length at $50 \%$ and $95 \%$ maturity) are estimated by fitting a binomial GLM on yearly bootstrapped maturity data with covariate length ( 500 resampled datasets). For more details, see Stock Annex.

## Size distribution resampling

The LBSPR model is fitted on 1000 bootstrapped size composition data and parameter sets. While input parameters were randomly generated/drawn as per Table 2.3.1, the generation of the randomized datasets is twofold:

1. random attribution of unclassified individuals between coastal and NEA cod, based on the size-based stock segregation model (section B.1) and using a binomial random generator: the number of coastal cod is drawn for each stratum defined by a combination size class, area, year, quarter and gear, based on the number of unclassified cod in the stratum and the probability P (coastal I size, area, year, quarter, gear) from the model described in section C.1.
2. bootstrap of the length composition within years: drawing the same number of individuals within each year of data from step 1, with replacement.
For each of the 1000 randomized data and parameter set, SPR, F/M and the selectivity parameter SL50\% and SL95\% are estimated and their resulting distributions evaluated.

### 2.3.6 Results of the Assessment (Figure 2.3.6-Figure 2.3.13)

### 2.3.6.1 Standardized CPUE index

The final standardized CPUE index for coastal cod indicates a general declining trend in all areas and quarter since 2007 with some interannual variability with a possible increase (large uncertainty) in 2020 (Figures 2.3.6 and 2.3.7).

The final standardized CPUE index for coastal cod indicates general stability since 2007 with some interannual variability and a possible increase (large uncertainty) in 2020. A declining trend is, however, seen in the southernmost part of the area, i.e. Møre-Trøndelag (statistical area 07).

A slightly new CPUE index of abundance was made as an extra check of the large uncertainty in 2020. Here we included the boat effect as a fixed effect since the model fit was much better than having the boat as a random effect, and then using one of the boats that was fishing for several years. This was made to possibly account for the unbalanced boat/gear use in the time-series. Even if it reduced the variance in 2020, we believe that the extra variance created by adding new boats and new fishing grounds to the time-series should not be disregarded. This issue will be further investigated until next year's assessment.

### 2.3.6.2 Effort and CPUE from official landings statistics

It has also been investigated whether official reported landings and measures of fishing effort in the sales note statistics can provide a CPUE index that can be used in assessment and practical management. If so, this will give a much larger material than just a few boats in the Coastal Reference Fleet that primarily sample biological data from the fisheries. On the other hand, a reference fleet CPUE is more controlled (e.g. with regards to technology creep and fishing behaviour) than a full fleet CPUE.

The number of sales notes has been shown to give an overestimation of the fishing effort since a trip can give several sales notes by splitting the entire trip catch into several sales, each with its own sales note. We have therefore come to the conclusion that a trip best can be described by combining the vessel's "Registration mark" in the sales note statistics with "Last catch date", and this we define as a trip and estimate effort according to.

| Vessel <br> size/Year | $\mathbf{2 0 1 8}$ | 2019 | 2020 |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Number of <br> trips | Landed round <br> weight ( $t$ ) | Number of <br> trips | Landed round <br> weight $(t)$ | Number of <br> trips | Landed round <br> weight ( $t$ ) |
| (blank) | 680 | 29 | 605 | 30 | 603 | 33 |
| $<11 \mathrm{~m}$ | 4203 | 229 | 3814 | 191 | 4311 | 298 |


| Vessel size/Year | 2018 |  | 2019 |  | 2020 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of trips | Landed round weight ( t ) | Number of trips | Landed round weight ( t ) | Number of trips | Landed round weight ( t ) |
| 11-14.99 m | 1107 | 129 | 1221 | 145 | 1125 | 114 |
| 15-20.99 m | 89 | 24 | 99 | 20 | 71 | 19 |
| 21-27.99 m | 3 | 2 | 1 | 1 | 32 | 15 |
| $>=28 \mathrm{~m}$ | 1 | 3 | 1 | 0 | 8 | 1 |

The text table above shows the number of trips and landings (round weight) per vessel length group for cod caught inside 12 nautical miles during the second half-year during 2018-2020, all gears. This shows that the vessel length groups $<11-14.99 \mathrm{~m}$ represented by the coastal reference fleet (ch. 2.2.6) are responsible for most of the effort and cod landings. The $9-15 \mathrm{~m}$ vessels in the reference fleet represent the gear and vessel size category responsible for about $60 \%$ of the total annual cod commercial catches in the area, and $88 \%$ of the effort (fishing trips) and $86 \%$ of cod catches in the second half of the year.
Figures 2.3.8 and 2.3.9 show the effort and CPUE from official landings statistics from 2007-2020. These data show a similar development of the CPUE as the more controlled and standardized reference fleet data do. These time-series can also be used by managers to adjust the number of trips as a measure of effort adjustment.

### 2.3.6.3 Stochastic LBSPR outputs and interpretation

SPR and F/M distributions per year are compared to their reference points. Between 2007-2019 for instance, the mean SPR fluctuates between 20 and $30 \%$, with an overall downward trend (Figure 2.3.10), which places it below the target values ( $30-40 \%$ ) and - at the end of the series just below the limit reference point $20 \%$, generally accepted in the absence of further information on the stock dynamics (ICES 2018; Prince et al., 2020; Mace and Sissenwine, 1993). The relative fishing mortality $\mathrm{F} / \mathrm{M}$ is estimated above the value which achieve long-term $\mathrm{SPR}=40 \%$, or the more usual proxy $\mathrm{F} / \mathrm{M}=1$ and follows an upward trend (Figure 2.3.11). The decrease in the spawning potential ratio is concomitant with a decline of the size-based indicators $\mathrm{L}_{\text {max }} 5 \%$ (the mean length of the largest $5 \%$ of individuals in the catch) and mean length in catch (Figure 2.3.12). These all together depict a somewhat depleted and worsening stock status.

In the absence of clear information on the stock-recruitment relationship, a more legitimate reference point cannot be estimated and even a SPR of $30 \%$ should be considered as a potentially non-precautionary level, and SPR=40\% preferred as Bmsy proxy (Clark, 2002; Hordyk et al., 2015a). In conformity with ICES guidelines (ICES, 2018) and commonly used SPR-based proxies (Prince et al., 2020; Mace and Sissenwine, 1993), the corresponding limit reference point (proxy for $\mathrm{Bl}_{\mathrm{lim}}=\mathrm{BMSY} / 2$ ) should be $\mathrm{SPR}=20 \%$.

A simulation function in the LBSPR package allows to estimate a $\mathrm{F} / \mathrm{M}$ which, at equilibrium and given the parameters, lead to a chosen SPR. The estimated F/M can therefore be compared to FSPR40\%/M (Figure 2.3.11) or other usual proxies.

### 2.3.6.4 Total mortality (Z) from catch curves

Since catch in numbers-at-age data is available for this stock (Tables 2.3.2 and 2.3.3) for a longer period (1994-2020) it is possible to estimate the total mortality from catch-curve analyses. The
assumptions usually made for catch-curve analysis are that (1) there are no errors in the estimation of age composition, (2) recruitment is constant or at least varies without trend over time, (3) Z is constant over time and across ages, and (4) above some determined age, all animals are equally available and vulnerable to the fishery and the sampling process. The catch-curve estimates a single total mortality rate for all years/ages that compose its synthetic cohort, and this total mortality estimate is generally similar to the average of the true total mortality rate.

With the available catch-at-age data it was possible to estimate the average total mortality of ages $5-14$ for the years 1994-2020. Note that Tables 2.3 .2 and 2.3 .3 only present data up to age group $10+$, but catch-at-age data were available to the AFWG up to age group 15+. Figure 2.3 .13 shows a very stable level of the total mortality during the entire time-series, varying without trend around the long-term average of $\mathrm{Z}=0.75$. With natural mortality of 0.23 (at L-infinity) this implies fishing mortality around 0.5.

### 2.3.7 Comments to the Assessment

The assessment is rather uncertain. The reasons for this include highly uncertain data for the recreational catch and uncertainty in the catch split between Northeast Arctic cod and coastal cod, although the CPUE series is calculated for the second half of the year to minimize the mixing of the two stocks in the dataseries. The assessment is also dependent on the representativeness of the coastal reference fleet's gillnet CPUE series. Gillnet is responsible for most of the catches, and the $9-15 \mathrm{~m}$ vessels in the reference fleet represent the gear and vessel size category responsible for about $60 \%$ of the total annual cod commercial catches in the area, and $88 \%$ of the effort (fishing trips) and $86 \%$ of cod catches in the second half of the year.

Since ICES lacks a framework for using LBSPR directly as a basis for quota advice, LBSPR and length-based analyses have been used as additional information to assess the need for a $20 \%$ buffer in the "2 over 3" rule, as recommended by the benchmark reviewers.

### 2.3.8 Reference points

No biological reference points are established except the SPR and F/M reference levels often referred to in literature. See section 2.3.6.1 above.

### 2.3.9 Catch scenarios for 2022

The ICES Guidance for completing single-stock advice for category 3 stocks was applied (ICES, 2012,2021 ). A composite standardized CPUE index from the coastal reference fleet ( $9-15 \mathrm{~m}$ vessel length) in coastal waters between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ during quarters 3 and 4, between 2007-2020, is used as index for the stock development. The advice is based on the ratio of the two latest index values (index A) with the three preceding values (index B), multiplied by the average catches for years 2018-2020 (Table 2.3.7-Table 2.3.8). The index is estimated to have increased by less than $20 \%$ and thus the uncertainty cap was not applied. Fishing pressure is thought to be above, and stock size is thought to be below, possible MSY reference points; therefore, the precautionary buffer was applied in the advice. Discarding (dead fish) is known to take place (less than $5 \%$ in the commercial fishery (Berg and Nedreaas 2021), and about 7\% in the rod and line sector of the recreational fishery), but ICES cannot quantify the corresponding catch.
The corresponding catch advice for 2022 is estimated to 7613 tonnes. Assuming recreational catches at 4202 tonnes, this implies a commercial catch of no more than 3411 tonnes. The catch advice is a decrease relative to the average catches 2018-2020 because of the application of the precautionary buffer, but an increase relative to the catch in 2020.

## Alternative 1 - Index values weighted with the inverse variance

Since the CPUE index for the stock development is calculated with variance, the AFWG did an alternative " 2 over 3 " estimation using indices A and B weighted by the inverse variance, especially since the last CPUE year (2020) had a relatively large variance. This gives an index ratio $\mathrm{A} / \mathrm{B}=1.029$ (Table 2.3.7) and corresponding catch advice for 2022 of 6666 tonnes when also using the $20 \%$ precautionary buffer.

## Alternative 2 - Using the rfb-rule (WKLIFE X)

ACOM intends to implement WKLIFE X methods (ICES 2020, Annex 3) in 2022. The AFWG was informed that a workplan will be developed for training, technical guidelines, special implementation workshops, and a big review group will be initiated later in 2021.

In this year's advice "season", ICES will hence provide advice using the "old" methods UNLESS a stock was benchmarked with the new WKLIFE X methods.

WKLIFE has developed a harvest control rule to provide MSY advice for category 3 stocks based on the " 2 over 3 rule". The recommended harvest rule, i.e. the rfb-rule, improves on the " 2 over $3^{\prime \prime}$ rule with the addition of multipliers based on the stock's life-history characteristics, the status of the stock in terms of relative biomass, and the status of the stock relative to a target reference length. The necessary parameters for using the rfb-rule were estimated during the benchmark assessment for this stock in February 2021 (WKBARFAR), and are presented in Tables 2.3.1 and 2.3.7. The corresponding catch advice will be higher than using the "old" " 2 over 3 rule".

### 2.3.10 Management considerations

Norwegian coastal cod is taken as part of a mixed fishery with Northeast Arctic cod (cod.27.1-2), from which it cannot be visually distinguished. Without the option of setting a direct TAC, the coastal cod stocks are managed by technical regulatory measures. Despite management actions, the previous management plan has not led to significantly reduced fishing mortality. A new plan is therefore required, with regulations better targeted to areas and seasons where catches of coastal cod are high. The split of the coastal cod stock in two units - one data rich in the north and one data poor in the south - combined with improved genetic stock identification techniques improves the spatial resolution of the assessment and allows development of more targeted management measures. The stock split follows the Norwegian catch reporting areas, with areas $0,3,4$, and 5 encompassing the northern stock, and areas 6 and 7 encompassing the southern (Figure 2.1).

The zero-catch advice for cod.27.1-2coastN (Northern Norwegian coastal cod) and non-zero catch advice for cod.27.2coastS (Southern Norwegian coastal cod) are not necessarily indicative of a better state for the southern stock. The difference is primarily due to the default ICES advice arising from the use of an analytic category 1 assessment in the north and a data-limited category 3 assessment in the south. Furthermore, the use of a longer time-series for the northern stock permits comparison with reference points from a higher stock state. Developing and adopting rebuilding plans for these two stocks should resolve this discrepancy.

ICES finds it difficult to give precise catch advice when the recreational catches, likely contributing more than $50 \%$ of total catches, are poorly estimated. A prerequisite for more accurate future assessments is a better estimation of the recreational catches.

### 2.3.11 Rebuilding plan for coastal cod

The Norwegian Ministry of Fisheries is working on a new rebuilding plan. Fisheries scientists need to discuss with managers, how to facilitate rebuilding of the stock, evaluate rebuilding targets and measures to avoid high fishing pressure in areas with high fractions of coastal cod. Stronger restrictions are required in all areas where coastal cod is distributed. Until a longer perspective rebuilding plan is established, the necessary management action for next year will be to reduce the fishery so that the combined commercial and recreational catches will become at least $6 \%$ lower than the three last years' average.

### 2.3.12 Recent ICES advice

For the years 2004-2011, the advice was; No catch should be taken from this stock and a recovery plan should be developed and implemented.
For 2012, and later the advice has been to follow the rebuilding plan. The latest ICES advice strongly recommends a new rebuilding plan.

### 2.3.13 Figures and tables

Table 2.3.2. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Estimated commercial landings in numbers ('000) at-age, and total tonnes by year.

|  | Age |  |  |  |  |  |  |  |  | Tonnes <br> Landed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |
| 1994 | 1 | 7 | 111 | 288 | 361 | 279 | 158 | 71 | 112 | 6381 |
| 1995 | 3 | 32 | 210 | 399 | 491 | 467 | 267 | 114 | 96 | 8936 |
| 1996 | 2 | 64 | 242 | 384 | 304 | 253 | 130 | 36 | 44 | 6207 |
| 1997 | 2 | 117 | 171 | 212 | 189 | 185 | 131 | 44 | 33 | 4746 |
| 1998 | 20 | 177 | 446 | 496 | 332 | 109 | 82 | 22 | 23 | 6200 |
| 1999 | 3 | 116 | 313 | 308 | 255 | 123 | 53 | 66 | 26 | 5522 |
| 2000 | 2 | 242 | 697 | 411 | 159 | 57 | 51 | 17 | 37 | 5838 |
| 2001 | 2 | 94 | 423 | 457 | 304 | 149 | 52 | 17 | 86 | 5250 |
| 2002 | 9 | 88 | 360 | 409 | 441 | 138 | 52 | 12 | 16 | 6937 |
| 2003 | 23 | 204 | 237 | 571 | 398 | 380 | 112 | 22 | 53 | 8905 |
| 2004 | 5 | 112 | 334 | 260 | 400 | 232 | 139 | 35 | 26 | 6866 |
| 2005 | 2 | 65 | 381 | 522 | 445 | 262 | 122 | 37 | 19 | 8005 |
| 2006 | 10 | 48 | 308 | 617 | 565 | 179 | 99 | 54 | 50 | 8612 |
| 2007 | 11 | 154 | 364 | 497 | 379 | 113 | 51 | 23 | 29 | 7695 |
| 2008 | 31 | 103 | 893 | 665 | 195 | 265 | 69 | 38 | 47 | 9889 |


|  | Age |  |  |  |  |  |  |  |  | Tonnes <br> Landed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |
| 2009 | 1 | 224 | 663 | 259 | 311 | 107 | 74 | 42 | 20 | 7145 |
| 2010 | 5 | 115 | 400 | 434 | 245 | 260 | 50 | 36 | 45 | 7634 |
| 2011 | 3 | 59 | 310 | 484 | 267 | 194 | 65 | 36 | 35 | 7128 |
| 2012 | 28 | 113 | 268 | 501 | 317 | 279 | 73 | 36 | 36 | 8187 |
| 2013 | 5 | 54 | 239 | 214 | 248 | 169 | 80 | 27 | 16 | 5131 |
| 2014 | 1 | 56 | 166 | 390 | 265 | 226 | 79 | 43 | 38 | 6244 |
| 2015 | 21 | 149 | 257 | 229 | 263 | 120 | 69 | 37 | 41 | 5004 |
| 2016 | 1 | 83 | 248 | 313 | 206 | 200 | 121 | 66 | 83 | 5962 |
| 2017 | 13 | 73 | 275 | 279 | 157 | 97 | 70 | 24 | 34 | 4159 |
| 2018 | 9 | 57 | 131 | 298 | 255 | 141 | 90 | 36 | 32 | 4436 |
| 2019 | 4 | 34 | 85 | 101 | 128 | 121 | 77 | 21 | 24 | 2965 |
| 2020 | 1 | 46 | 164 | 140 | 144 | 79 | 84 | 37 | 16 | 3481 |

Table 2.3.3. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Total estimated catch number ('000) at age, including recreational and tourist catches.

|  | Age |  |  |  |  |  |  |  |  | Tonnes <br> landed | Hereof <br> rec. ( t ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |  |
| 1994 | 2 | 14 | 207 | 538 | 676 | 523 | 296 | 132 | 210 | 11937 | 5556 |
| 1995 | 4 | 51 | 341 | 647 | 797 | 757 | 433 | 184 | 155 | 14492 | 5556 |
| 1996 | 3 | 120 | 455 | 723 | 572 | 476 | 245 | 68 | 82 | 11687 | 5480 |
| 1997 | 5 | 253 | 369 | 456 | 407 | 399 | 283 | 95 | 72 | 10226 | 5480 |
| 1998 | 38 | 334 | 842 | 937 | 628 | 207 | 155 | 42 | 43 | 11718 | 5518 |
| 1999 | 5 | 226 | 610 | 600 | 497 | 240 | 103 | 128 | 51 | 10776 | 5254 |
| 2000 | 3 | 456 | 1311 | 773 | 299 | 107 | 96 | 32 | 69 | 10979 | 5140 |
| 2001 | 3 | 184 | 832 | 897 | 598 | 293 | 101 | 34 | 169 | 10315 | 5065 |
| 2002 | 15 | 153 | 627 | 711 | 768 | 240 | 91 | 22 | 28 | 12077 | 5140 |
| 2003 | 36 | 325 | 377 | 907 | 633 | 605 | 178 | 35 | 85 | 14159 | 5254 |
| 2004 | 9 | 194 | 581 | 451 | 695 | 403 | 242 | 60 | 45 | 11931 | 5065 |
| 2005 | 3 | 105 | 619 | 848 | 722 | 426 | 197 | 61 | 31 | 12994 | 4989 |
| 2006 | 16 | 76 | 484 | 968 | 888 | 282 | 156 | 84 | 79 | 13525 | 4913 |


|  | Age |  |  |  |  |  |  |  |  | Tonnes <br> landed | Hereof <br> rec. ( t ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |  |
| 2007 | 18 | 252 | 597 | 814 | 620 | 185 | 83 | 38 | 47 | 12609 | 4913 |
| 2008 | 46 | 153 | 1330 | 990 | 290 | 395 | 103 | 56 | 71 | 14727 | 4838 |
| 2009 | 1 | 375 | 1109 | 433 | 519 | 178 | 124 | 70 | 34 | 11945 | 4800 |
| 2010 | 7 | 187 | 651 | 706 | 398 | 423 | 81 | 58 | 74 | 12434 | 4800 |
| 2011 | 5 | 98 | 518 | 811 | 447 | 325 | 109 | 59 | 58 | 11928 | 4800 |
| 2012 | 45 | 179 | 425 | 795 | 502 | 442 | 115 | 57 | 58 | 12987 | 4800 |
| 2013 | 9 | 105 | 463 | 414 | 480 | 327 | 154 | 52 | 31 | 9931 | 4800 |
| 2014 | 1 | 100 | 293 | 690 | 469 | 400 | 140 | 76 | 68 | 11044 | 4800 |
| 2015 | 41 | 293 | 503 | 449 | 515 | 234 | 135 | 72 | 80 | 9804 | 4800 |
| 2016 | 2 | 151 | 448 | 566 | 371 | 360 | 218 | 120 | 150 | 10762 | 4800 |
| 2017 | 28 | 158 | 592 | 600 | 337 | 208 | 152 | 51 | 73 | 8959 | 4800 |
| 2018 | 19 | 118 | 272 | 620 | 532 | 293 | 187 | 75 | 66 | 9236 | 4800 |
| 2019 | 12 | 88 | 223 | 265 | 336 | 316 | 201 | 54 | 63 | 7765 | 4800 |
| 2020 | 1 | 97 | 342 | 293 | 301 | 166 | 177 | 78 | 34 | 7287 | 3806 |

Table 2.3.4. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Commercial catch in 2020 by gear and Norwegian statistical fishing area. Both fishing areas lie within ICES Division 2.a.

| Year | 2020 |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Area | 06 | 07 | Total between 62 and $67^{\circ} \mathrm{N}$ | 67.3 |
| Gillnet | 1355 | 988 | 2343 | 3.1 |
| Longline/Handline |  |  | 29.6 |  |
| Danish seine | 14 | 93 | 107 |  |
| Trawl | 366 | 665 | 1031 | 3481 |
| Others* | 1735 | 1746 |  |  |
| Total |  |  |  |  |

*in 2020, longline, handline and Danish seine are all included in the 'others' category.

Table 2.3.5. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Mean weight at age in the catch.

| CWT | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1994 | 1.028 | 1.537 | 2.206 | 2.985 | 3.822 | 4.908 | 5.954 | 7.468 | 9.571 |


| CWT | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 0.845 | 1.392 | 1.950 | 2.603 | 3.649 | 4.811 | 6.076 | 7.404 | 10.566 |
| 1996 | 1.177 | 1.975 | 2.554 | 3.392 | 4.186 | 5.242 | 6.429 | 7.283 | 11.591 |
| 1997 | 1.348 | 2.004 | 2.611 | 3.439 | 4.282 | 5.387 | 6.563 | 7.467 | 10.828 |
| 1998 | 1.007 | 1.737 | 2.454 | 3.373 | 4.483 | 5.484 | 6.914 | 7.825 | 14.092 |
| 1999 | 1.459 | 2.231 | 2.927 | 3.800 | 4.854 | 6.032 | 7.009 | 8.257 | 12.088 |
| 2000 | 1.344 | 1.971 | 2.811 | 3.568 | 4.610 | 5.588 | 6.860 | 7.815 | 11.806 |
| 2001 | 0.565 | 0.981 | 1.533 | 2.250 | 3.129 | 4.160 | 5.375 | 6.722 | 16.118 |
| 2002 | 1.372 | 2.330 | 3.302 | 4.199 | 5.225 | 6.290 | 7.226 | 9.768 | 13.031 |
| 2003 | 1.312 | 2.143 | 2.962 | 3.899 | 4.702 | 5.648 | 6.616 | 7.425 | 11.376 |
| 2004 | 1.368 | 2.124 | 2.758 | 3.684 | 4.705 | 5.858 | 6.874 | 7.901 | 11.117 |
| 2005 | 1.488 | 2.332 | 2.990 | 3.701 | 4.562 | 5.637 | 6.699 | 7.703 | 10.364 |
| 2006 | 1.526 | 2.158 | 2.866 | 3.790 | 4.703 | 5.769 | 6.725 | 7.876 | 10.103 |
| 2007 | 1.613 | 2.295 | 3.285 | 4.337 | 5.744 | 7.105 | 8.397 | 9.991 | 12.359 |
| 2008 | 1.455 | 2.221 | 3.179 | 3.932 | 5.443 | 6.533 | 7.990 | 8.341 | 11.107 |
| 2009 | 1.667 | 2.135 | 3.234 | 4.207 | 5.279 | 6.527 | 7.568 | 7.606 | 11.305 |
| 2010 | 1.480 | 2.262 | 3.325 | 4.431 | 5.534 | 6.335 | 7.598 | 9.048 | 9.543 |
| 2011 | 1.381 | 2.127 | 3.172 | 4.263 | 5.511 | 6.510 | 8.012 | 9.032 | 11.065 |
| 2012 | 1.214 | 2.012 | 3.011 | 4.302 | 5.520 | 6.686 | 8.188 | 9.569 | 11.635 |
| 2013 | 1.269 | 2.027 | 3.092 | 4.024 | 5.268 | 6.370 | 7.524 | 8.918 | 12.241 |
| 2014 | 1.304 | 2.194 | 3.047 | 3.998 | 4.959 | 6.115 | 7.181 | 8.234 | 11.537 |
| 2015 | 1.219 | 1.832 | 2.726 | 3.797 | 4.627 | 5.845 | 7.009 | 8.195 | 10.981 |
| 2016 | 1.339 | 1.930 | 2.617 | 3.578 | 4.471 | 5.421 | 6.429 | 7.445 | 9.132 |
| 2017 | 1.529 | 2.022 | 2.750 | 3.663 | 4.543 | 5.612 | 6.542 | 7.489 | 9.678 |
| 2018 | 1.190 | 1.848 | 2.547 | 3.434 | 4.265 | 5.301 | 6.375 | 7.333 | 9.393 |
| 2019 | 1.662 | 2.283 | 3.120 | 3.895 | 4.840 | 5.796 | 6.743 | 7.737 | 9.548 |
| 2020 | 1.660 | 2.395 | 3.150 | 3.922 | 4.707 | 5.505 | 6.313 | 7.130 | 8.993 |

Table 2.3.6. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Composite standardized CPUE index from the coastal reference fleet during quarters 3 and 4, between 2007-2020. 95\% confidence interval (calculated using the approximation: mean +/-SD).

| Year | CPUE index | SD +/- |
| :---: | :---: | :---: |
| 2007 | 0.24 | 0.66 |
| 2008 | 0.38 | 0.89 |
| 2009 | 0.23 | 0.50 |
| 2010 | 0.14 | 0.32 |
| 2011 | 0.21 | 0.54 |
| 2012 | 0.18 | 0.49 |
| 2013 | 0.05 | 0.11 |
| 2014 | 0.12 | 0.27 |
| 2015 | 0.22 | 0.51 |
| 2016 | 0.24 | 0.54 |
| 2017 | 0.27 | 0.72 |
| 2018 | 0.11 | 0.28 |
| 2019 | 0.13 | 0.33 |
| 2020 | 0.35 | 0.96 |

Table 2.3.7. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Parameters used for calculating "2 over 3" and the "rfb" (ICES WKLIFE X 2021).
\(\left.\begin{array}{lll}\hline Parameter \& Value \& Value multiplied with <br>

precautionary buffer=0.8\end{array}\right]\)| Average CPUE 2019-2020 | 0.243 |  |
| :--- | :--- | :--- |
| Average CPUE 2016-2018 | 0.207 | 0.94 |
| Average CPUE 2019-2020 (weighted) | 0.154 | 0.82 |
| $r$ (plain) | 1.174 |  |
| $r$ (weighted) | 1.029 |  |
| Mean length in observed catch, $\mathrm{L}_{\mathrm{y}-1(2020)}$ | 73.7 cm |  |
| Length at modal abundance | 74 cm |  |
| $L c$ is defined as length at $50 \%$ of modal abundance | 61 cm |  |
| Linf | 95.45 cm |  |

\(\left.\begin{array}{lll}\hline Parameter \& Value \& Value multiplied with <br>

precautionary buffer=0.8\end{array}\right]\)| $\mathrm{L}_{\mathrm{F}=\mathrm{M}}=0.75 \mathrm{~L}_{\mathrm{c}}+0.25 \mathrm{~L}_{\mathrm{inf}}$ | 69.63 cm |
| :--- | :--- |
| $\mathrm{f}=\mathrm{L}_{\mathrm{y}-1} / \mathrm{L}_{\mathrm{F}=\mathrm{M}}$ | 1.06 |
| $\mathrm{I}_{\mathrm{y}-1}$ | 0.36 |
| $\mathrm{I}_{\mathrm{trigger}}=1.4 \mathrm{I}_{\text {loss }}$ | 0.07 |
| b | 1.0 |
| m when k=0.248 | 0.8 |
| Total factor rfbm (with plain r) | 1.00 |
| Total factor rfbm (with "weighted" r) | 0.87 |

Table 2.3.8. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. The basis for the catch scenarios ${ }^{\wedge}$.

| Index A (2019-2020) |  | 0.243 |
| :---: | :---: | :---: |
| Index B (2016-2018) |  | 0.207 |
| Index ratio (A/B) |  | 1.174 |
| Uncertainty cap | Not applied |  |
| Average catches for 2018-2020 |  | 8096 |
| Discard rate |  | Not quantified |
| Precautionary buffer | Applied | 0.8 |
| Catch advice * |  | 7613 |
| \% Advice change ** |  | -6\% |

${ }^{\wedge}$ The figures in the table are rounded. Calculations were done with unrounded inputs and computed values may not match exactly when calculated using the rounded figures in the table.

* [average catches for 2018-2020] $\times$ [index ratio] $\times$ [precautionary buffer].
** Advice value for 2022 relative to average catches for 2018-2020.


Figure 2.3.1. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Commercial and recreational catches. Recreational catches are fixed from 2009-2019 at 4800 tonnes.


Figure 2.3.2. Estimated landings of Northeast Arctic cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$.


Figure 2.3.3. Residual diagnostic plots for the final binomial model to differentiate coastal cod vs. NEAC. The panel on the left is a standard output from the residual diagnostics using the $R$ package DHARMa. The panel on the right plots the model standardized residuals against available covariates. Both panels indicate no significant issues with the final model.


Figure 2.3.4. Predicted probability of catching coastal cod based on the quarter (vertical panels), areas (horizontal panels), and years ( $x$-axis within each panel). The grey shaded polygon represents the $95 \%$ confidence interval.


Figure 2.3.5. Residual diagnostic plots for the final CPUE model fitted to cod data in area 6 and 7, and quarters 3 and 4. The top panel is the normal QQ-plot. The panel on the left is a standard output from the residual diagnostics using the $R$ package DHARMa. The panel on the right plots the model standardized residuals against available covariates. All panels indicate no significant (though some) issues with the final model.


Figure 2.3.6. Standardized CPUE index for coastal cod in area 6 and 7 during quarters 3 and 4, between 2007-2020. The grey shaded polygon represents the $95 \%$ confidence interval (calculated using the approximation mean $+/-1.96$ std which is why some values goes below 0 ).


Figure 2.3.7. Composite standardized CPUE index for coastal cod in area 6 and 7 during quarters 3 and 4, between 20072020. 95\% confidence interval (calculated using the approximation: mean +/-1.96 std.; negative values are therefore introduced in the plot as an artifact of this procedure) are given by error bars.

$$
\text { Distinct count of trips } \mathrm{N}_{\mathrm{r}} \text {, only length groups } 2 \text { and } 3 \text {, linear } y \text {-axis scaling }
$$



Figure 2.3.8. Fishing effort presented as the number of sales note trips for two boat sizes, LG2 = <11 m and LG3 = 1114.99 m , for areas $62-67^{\circ} \mathrm{N}$ in the second half of the year. Left panel: all gears; right panel: gillnet only. Note different y axes.


Figure 2.3.9. CPUE (kg cod per sales note trip) per boat size (LG1-LG6) for area $62-67^{\circ} \mathrm{N}$ in the 2 nd half of the year. Left panel: all gears; right panel: gillnet only.


Figure 2.3.10. Estimated spawning potential ratio (SPR) per year for coastal cod south of $67^{\circ} \mathrm{N}$. Mean (solid line) and confidence intervals (shaded red area, 95\% IQR), based on the stochastic LBSPR. The grey shaded area delimits the SPR30\%-40\% zone (common targets) and the dotted horizontal line the SPR20\% limit reference point (Prince et al., 2020).


Figure 2.3.11. Estimated fishing mortality, relative to natural mortality ( $F / M$ ) per year for coastal cod south of $67^{\circ} N$. Mean (solid line) and confidence intervals (shaded red area, 95\% IQR), based on the stochastic LBSPR.


Figure 2.3.12. Variations in time of the size-based indicators Lmax5\% and mean length in catch ( $\bar{L}$ ), and their reference points (mean and $95 \% \mathrm{CI}$ ). The reference points were estimated using the LBSPR simulation model together with the stochastic parameters detailed in Table 2.3 .1 (mortality scenario following Lorenzen, 1996) and SPRs of 40\% and 100\% (unfished).


Figure 2.3.13. Total mortality $(Z)$ estimated from catch curves (average over ages 5-14 in commercial and recreational catches) 1994-2020.


[^0]:    ICES INTERNATIONAL COUNCIL FOR THE EXPLORATION OF THE SEA
    CIEM CONSEIL INTERNATIONAL POUR L'EXPLORATION DE LA MER

