8 Blue shark in the North Atlantic (North of 5°N)

8.1 Stock distribution

There is a discrete North Atlantic stock of blue shark *Prionace glauca* (Heessen, 2003; Fitzmaurice *et al.*, 2005; ICCAT, 2008), with 5°N latitude as the southern stock boundary, and a separate South Atlantic stock (ICCAT, 2008). This delineation is based on mark-recapture data and oceanographic features. In addition, this division facilitates comparison with fisheries statistics of other North Atlantic stocks, such as tuna-like species, that have the same southern stock boundary. Hence, the ICES area is only part of the stock area.

Recent genetic studies on blue shark reveal genetic homogeneity across whole ocean basins in Atlantic (Verissimo *et al.*, 2017) and Pacific oceans (Ovenden *et al.*, 2009; Taguchi *et al.*, 2015). These are at odds with the currently assumed distinction of northern and southern stocks within each ocean basin. The bulk of the evidence gathered thus far indicates that the blue shark exhibits dispersal with gene flow over very large spatial scales, and little to no philopatry to the sampled nursery areas or to distinct ocean basins. However, in cases as in blue sharks where effective populations sizes are ~1000s, the levels of genetic divergence associated with migration rates which could lead to demographic connectivity (~10%; Hastings, 1993) may be difficult to detect using traditional molecular markers. In these cases, the precautionary approach in conservation and fisheries management would be to consider each nursery area as independent, with potentially different demographic parameters and vulnerability to fishing pressure. If each nursery area currently exchanges only a few migrant individuals per generation with other nurseries, the replenishment of each stock would be mostly dependent on recruit survival rather than on immigration from adjacent stocks.

8.2 The fishery

8.2.1 History of the fishery

In recent years, more information has become available about fisheries taking blue shark in the North Atlantic. Although available data are incomplete, they offer information on the situation in fisheries and trends. There are no large-scale target fisheries for blue shark, but it is a major bycatch in tuna and billfish fisheries, where it can comprise up to 70% of the total catches and even exceed the catch of target species (ICCAT, 2005). In the North Atlantic, EU fleets (Portugal and Spain) are responsible for approximately 82% of the total landings (Anon., 2015).

Observer data indicates that substantially more blue sharks are caught as bycatch than reported in landings statistics. Blue sharks are also caught, in considerable numbers, in recreational fisheries, including in the ICES area (Campana *et al.*, 2005).

Since 1998, there has been a Basque artisanal longline fishery targeting blue shark and other pelagic sharks in the Bay of Biscay from June to November (Díez *et al.*, 2007). Initially 3–5 vessels were involved but, as a consequence of changes in local fishing regulations, the number of vessels has reduced to two after 2008.

In the North Atlantic, thirteen fisheries (in descending order of volume: EU-Spain, EU-Portugal, Japan, Canada, USA_LL, Chinese Taipei, EU-France, Belize, Panama, USA_SP., China PR, Korea and, Venezuela) accounted for 99% of the total removals (1990–2014). The majority (except: USA

sport fishery, EU-France unclassified gear) are longline fisheries (Anon., 2015). There are also blue shark landings in Mediterranean fisheries (Anon., 2015).

In 2015, prior to their most recent stock assessment, ICCAT nominal catch statistics of blue shark (by stock, flag and gear) were reviewed. No major updates were made to the historical catch series, and only recent years of official catches were updated. Before 1997, there was a lack of official catch statistics for some of the main fishing nations operating in the stock area.

8.2.2 The fishery in 2020

No new information.

No major changes noted in 2020, although potential changes to fishing effort caused by the effects of COVID-19 have not yet been quantified.

8.2.3 Advice applicable

ACOM has never provided advice for blue shark in the ICES area. Assessment of this stock is considered to be the responsibility of ICCAT. In July 2015, members of WGEF participated in the ICCAT blue shark stock assessment meeting that took place in Lisbon, Portugal (ICCAT, 2015).

In 2015, ICCAT considered that the status of the North Atlantic stock is unlikely to be either overfished or subject to overfishing. However, due to the level of uncertainty in the assessment results no specific management recommendations were provided (ICCAT, 2015).

ICCAT adopted Rec. 16-12, which in paragraph 2 establishes a catch limit for blue sharks in the North Atlantic (39,102 t as the average of two consecutive years). This measure came into being in 2017 and the Standing Committee on Research and Statistics (SCRS) has been in a position to assess the effect of this measure as yet. However, SCRS data show that preliminary catches in 2017 and 2018 were 39 675 t and 33 853 t, respectively (SCRS, 2019).

In 2019, these catch limits were further refined when ICCAT adopted Rec. 19-07. This states that: "An annual TAC of 39 102 t for North Atlantic blue shark is established. The annual TAC may be revised subject to a decision of the Commission based on the updated advice of the SCRS in 2021, or at an earlier stage if enough information is provided by the SCRS". Catch limits for the EU, Japan and Morocco were set at 32 578 t, 4 010 t and 1 644 t, respectively, with all other CPCs to "endeavour to maintain their catches at recent levels".

8.2.4 Management applicable

There are no measures fully regulating all the catches of blue shark in the North Atlantic.

European regulations for annual fishing opportunities have given an overall TAC (39 102 t) for blue shark in the Atlantic Ocean north of 5°N since 2017. Whilst this nominal TAC has remained unchanged, an allocation key was included in the 2020 fishing opportunities (Council Regulation (EU) 2020/123), under which the EU quota was set at 32 578 t (83.3% of the 39 102 t TAC, and in accordance with ICCAT Rec. 19-07), and this was allotted to Spain (27 062 t), Portugal (5 363 t), France (152 t) and Ireland (1 t).

EC Regulation No. 1185/2003 (updated by EU Regulation No 605/2013) prohibits the removal of shark fins of these species, and subsequent discarding of the body. This regulation is binding on EC vessels in all waters and non-EC vessels in Community waters.

8.3 Catch data

8.3.1 Landings

It is difficult to accurately quantify landings of blue shark in the North Atlantic. Data are incomplete, and the generic reporting of shark catches has resulted in underestimations. Landing data from different sources (ICCAT, FAO and national statistics) can vary (Figures 8.1–8.3). Table 8.1 gives the catch data (total landings and discards by stock, flag and major gears) collated by IC-CAT, which appears to provide the most complete catch data for this stock, though there can be small changes in these data over time (Tables 8.2–8.3). ICCAT considers that reported landings of blue shark were underestimated in the early part of the time-series (prior to 1997), with official landings and estimates of a comparable magnitude since 1997, when annual landings have been *ca.* 20 000–40 000 t. In the North Atlantic, blue shark is reported predominantly by Spain, Portugal, Japan, USA and Canada (Figure 8.1).

In 2015, alternative approaches to estimate catch series were discussed by ICCAT (Anon., 2015), including (i) ratios between blue shark catches and species-specific catches derived from ICCAT Task I data; (ii) catch/effort and standardised CPUE; and (iii) shark fin trade data. Figure 8.4 shows the catch series (1971–2013) for North Atlantic blue shark available for the 2015 stock assessment (SA2015), the 2008 stock assessment catches (SA2008), and the catch series obtained using shark-fin ratios (three different series, see for example Clarke et al., 2006). Both stock assessment series followed a similar trend (but with large differences in some years) with catches oscillating several times between 15 000 t and 55 000 t. The three shark-fin series showed completely different trends (continuous upward trend) with catches starting around 10 000 t in the 1980s and growing to nearly 60 000 t in 2011 (Anon., 2015). Generally, the overall data for blue shark (and sharks in general) reported to ICCAT has improved over time (more complete series by species, lesser quantities of unclassified sharks, less weight of unclassified gears in the shark series, etc.). However, many unclassified shark species, mostly grouped by family (e.g. Lamnidae, Carcharhinidae, Sphyrnidae) and genera (e.g. Rhizoprionodon, Carcharhinus, Sphyrna and Alopias spp.) were reported to ICCAT in the past. The largest portion of unclassified sharks (1982-2013) is concentrated in longline and gillnet fisheries (Anon., 2015).

Japanese catches (landings and discards) from tuna longliners in the North Atlantic are estimated to have fluctuated between 1 400–2 400 t in 2006–2014, but a large increase to about 8 200 t is observed in 2015. These are higher than reported landings of the target species (bluefin tuna) from Japanese longliners in this period (ICCAT, 2008). Another study of Japanese bluefin tuna longline fisheries showed that the ratio of blue shark to the target species was about 1:1 (Boyd, 2008). Data from observations onboard a Chinese Taipei (Taiwanese) vessel targeting bluefin tuna in the southern North Atlantic showed that blue shark accounted for 76% of shark bycatch, though no information was presented on the percentage of blue shark in the total catch (Dai and Jang, 2008). Together, blue shark and shortfin mako account for between 69% and 72% of catches from Spanish and Portuguese surface longliners in the North Atlantic (Oceana, 2008).

The most recent ICCAT data publicly available for Task 1 data (landings and dead discards) for blue shark from the North Atlantic indicated a decrease in 2017–2019 (though 2019 data may be incomplete) compared to 2016 (Table 8.3). This would imply a reduction in landings in line with the ICCAT Recommendations relating to catch limits (see Section 8.2.3).

8.3.2 Discards

Historically, the relative low value of blue shark meant that it was not always retained for the market, with the fins the most valuable body part. In some fisheries the fins were retained and the carcasses discarded. In 2013, EU prohibited this practice (see Section 8.2.4).

Accurate estimates of discarding are required to quantify total removals from the stock. Currently no such estimates are available. Differences between estimated and reported catch in various fisheries (ICCAT, 2008 and references cited therein) suggest that discarding is widespread in fisheries taking blue shark.

Discard estimates are available for fisheries from Chinese Taipei, Korea Rep., USA, and UK (Bermuda) in recent years and from 2000 onwards from USA. However, they represent a limited part of total discards. The full extent of blue shark bycatch cannot be assessed using the data available, but evidence suggests that longline operations can catch more blue sharks than target species. There is considerable bycatch of blue sharks in Japanese and Taiwanese tuna longliners operating in the Atlantic. However, it is not possible, to estimate discard rates from these fleets from the information available. Discards can be assumed to be far higher than reported (Campana *et al.*, 2005), especially in high seas fisheries.

Information on elasmobranchs discards in demersal otter trawl, deep-water set longlines, set gillnet and trammel net fisheries for ICES Division 9.a (2004–2013) showed that blue shark was caught infrequently and discarded in the longline fishery but not in the other fisheries (Prista *et al.*, 2014).

8.3.3 Discard survival

Blue shark is one of the most frequent shark species captured in pelagic longline fisheries, and there are several estimates of survival (Boggs, 1992; Francis et al., 2001; Campana et al., 2005; Diez and Serafy, 2005). It is thought that most discards of whole sharks would be alive on return to the sea. For instance, discard survival rate is estimated to be about 60% in longline fisheries and 80% in rod and reel fisheries (Campana et al., 2005). More generally, the at-vessel mortality of longline-caught blue shark ranges from about 5–35% (summarised in Ellis et al., 2014 WD). Discard survival in such fisheries can be influenced by several factors, including hook type, soak time and size of shark. However, discarding can increase overall mortality attributable to fisheries: a study conducted on Canadian pelagic longliners targeting swordfish in the Northwest Atlantic (Campana et al., 2009) showed that "overall blue shark bycatch mortality in the pelagic longline fishery was estimated at 35%, while the estimated discard mortality for sharks that were released alive was 19%. The annual blue shark catch in the North Atlantic was estimated at about 84 000 t, of which 57 000 t is discarded. A preliminary estimate of 20 000 t of annual dead discards for North Atlantic blue sharks is similar to that of the reported nominal catch, and could substantially change the perception of population health if incorporated into a population-level stock assessment". The survival rate at hauling for blue shark was estimated to be 49% for the French pelagic longliners targeting swordfish in the southwest Indian Ocean. Field trials conducted with gears equipped with hook timers indicated that 29% were alive 8 h after their capture (Poisson et al., 2010). The survival rate of blue shark (at haul back) after a night-time soak may be lower than that during day-time soaks.

8.3.4 Quality of catch data

Catch data are incomplete, and the extent of finning in high seas fisheries is unclear. The historical use of generic shark categories is also problematic, although European countries now report more species-specific data. In 2012, the ICCAT Secretariat noted some large discrepancies between the data in the EURO-STAT database and that of the ICCAT database, with EUROSTAT records showing captures almost double those of ICCAT in recent years.

Methods developed to identify shark species from fins (Sebastian *et al.*, 2008; Holmes *et al.*, 2009) could help to gather data on species targeted by illegal fishers, this information will greatly assist in management and conservation.

The variability of blue shark mortality estimates, relating to the proportion of live discards, hampers the estimation of total removals, although there are improving approaches to reporting of live discards to the ICCAT SCRS (Anon., 2015).

Given the uncertainty on the 2015 assessment of blue shark North Atlantic stock, ICCAT recommended continued monitoring of the fisheries by observer and port sampling programmes (IC-CAT, 2015).

8.4 Commercial catch composition

No new information.

8.4.1 Conversion factors

Information on the length–weight relationship is available from several scientific studies (Table 8.4), as are the relationships between various length measurements (Table 8.5a and 8.5b). Campana *et al.*, 2005 calculated the conversion relationships between dressed weight (W_D) and live weight or round weight (W_R) for NW Atlantic blue shark (n = 17) to be $W_R = 0.4 + 1.22 W_D$ and $W_D = 0.2 + 0.81 W_R$.

For French fisheries, the proportion of gutted fish to round weight is 75.19%. There is also a factor for landed round weight to live weight (96.15%), meaning that there is a 4% reduction in weight because of lost moisture (Hareide *et al.*, 2007). Various estimates of fin weight to body weight are available (Mejuto and García-Cortés, 2004; Santos and Garcia, 2005; Hareide *et al.*, 2007; Santana-Garcon *et al.*, 2012; Biery and Pauly, 2012).

8.5 Commercial catch and effort data

For the North Atlantic stock, reported catches showed an increase in 1998, followed by a gradual decline until 2002 and then an increase (Figure 8.3). The CPUE input data available were comprehensively described and presented in the 2015 blue shark data preparatory meeting report (Anon., 2015). Following the work conducted for the 2008 SCRS blue shark stock assessment, CPUE were combined through a GLM with two choices of weighting: by the catch of the flag represented by each index and by the area of the flag represented by each index. Additionally, a hierarchical index of abundance that combines all available indices into a single series was also developed. However, it was noted that the process of combining CPUE indices was discouraged as they tend to mask the individual trends of the series and the underlying reasons as to why the series are different. It also indicated that some models can stochastically make use of the different series without need to combine these indices. It was suggested that it may be more useful to group CPUEs according to similar trends, and to include these as separate scenarios as was discussed during the 2015 bigeye tuna assessment.

Table 8.6 shows the various CPUE indices currently available (EU-Portugal, EU-Spain, USA, Japan, Chinese Taipei, and Venezuela), which have been considered for use in the assessment.

These CPUE indices show a relatively flat trend throughout the time-series, but with high variance (Tables 8.6–8.7; Figure 8.5).

8.6 Fishery-independent surveys

No fishery-independent data are available for the NE Atlantic, although such data exist for parts of the NW Atlantic (Hueter *et al.*, 2008). A survey from 1977–1994 conducted by the US NMFS documented a decline among juvenile male blue sharks by 80%, but not among juvenile females, which also occur in fewer numbers in the area, the western North Atlantic off the coast of Massachusetts (Hueter *et al.*, 2008). The authors concluded that vulnerability to overfishing in blue sharks is present despite their enhanced levels of fecundity relative to other carcharhinid sharks.

8.7 Life-history information

Blue shark has one of the widest ranges of all the shark species, being common in pelagic, oceanic waters in tropical and temperate oceans worldwide, as well as closer to shore (Coelho *et al.*, 2018). Various papers have reviewed the biology of blue shark (Nakano & Seki, 2003; da Silva *et al.*, 2021).

In a satellite telemetry study, Queiroz *et al.* (2010) described complex and diverse types of behaviour depending on water stratification and/or depth (Figure 8.6). Females tagged in the Western channel were able to spend up to 70 days in the shelf edge area in the Bay of Biscay; whereas tagged juveniles showed relatively extensive vertical movements away from the southern nursery areas. Results indicated that the species inhabited waters with a wide temperature range (10–20°C).

The US National Marine Fisheries Service also conducts a Cooperative Shark Tagging Programme (CSTP; Kohler *et al.*, 1998; NMFS, 2006), with tagging in the NE Atlantic also being undertaken under the auspices of the Inshore Fisheries Ireland (formerly the Irish Central Fishing Board) Tagging Programme (Green, 2007 WD) and UK Shark Tagging Programme, and there have been other earlier European tagging studies (e.g. Stevens, 1976). The tag and release results presented by ICCAT (2012; Figure 8.7) highlights the large number of blue shark tagged to date, and the extensive horizontal movements undertaken by blue shark in the Atlantic.

In Australian waters, blue shark exhibits oscillatory dive behaviour between the surface layers to as deep as 560–1000 m. Blue sharks mainly occupied waters of 17.5–20.0°C and spent 35–58% of their time in <50 m depths and 10–16% of their time >300 m (Stevens *et al.*, 2010). The distribution and movements of blue sharks are strongly influenced by seasonal variations in water temperature, reproductive condition and prey availability. Blue shark often occurs in large single-sex schools containing individuals of similar size.

Adult blue sharks have no known predators, although sub-adults and juveniles are eaten by shortfin mako, white shark and sea lions. Fishing is likely to be a major contributor to adult mortality. An estimation of fishing mortality rate via satellite tagged sharks being recaptured by fishing vessels ranged from 9–33% (Queiroz *et al.*, 2010).

Various studies have compiled biological information on this species in the North Atlantic and other areas, with some of these data summarized in Tables 8.4 (length–weight relationships), 8.3a and 8.3b (length–length relationships), Table 8.8 (growth parameters) and Table 8.9 (other life-history parameters). Based on life-history information, the blue shark is considered to be among the most productive shark species (ICCAT, 2008).

New life history inputs were obtained from data first assembled at the ICCAT 2014 Intersessional Meeting of the Shark Species Group (SCRS/2014/012) and additional information provided

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during the 2015 blue shark data preparatory meeting (SCRS/2015/142). These included maximum population growth rates (rmax) and steepness (h) values of the Beverton–Holt stock–recruitment relationship for North and South Atlantic stocks of blue shark, based on the latest biological information available gathered at the 2015 blue shark data preparatory meeting. To encompass a plausible range of values, uncertainty in the estimates of life history inputs (reproductive age, lifespan, fecundity, von Bertalanffy growth parameters, and natural mortality) was incorporated through Monte Carlo simulation by assigning statistical distributions to those biological traits in a Leslie matrix approach. Estimated productivity was high (rmax = 0.31-0.44 y⁻¹ for the North Atlantic stock), similar to other stocks of this species. Consequently, analytically derived values of steepness were also high (h = 0.73-0.93 for the North Atlantic stock).

The influence of different biological parameters (e.g. growth coefficients, reproductive periodicity, first maturation age, natural mortality and longevity) on estimated blue shark productivity was assessed. Age at first maturity and growth coefficients substantially influenced the estimated productivity (e.g. a low age at first maturity and high growth coefficient results in high productivity), and reproductive periodicity also affected productivity (i.e. a longer breeding period decreased productivity). Biological parameters should be carefully considered when they are used in the stock analysis, especially when estimated productivity is inconsistent with trends in abundance indices. The level of depletion experienced by blue shark stocks may affect the productivity or population growth through density dependence, and differences in environmental water temperature may also affect growth rates (Anon., 2015).

8.8 Exploratory assessment models

8.8.1 Previous assessments

In 2004, ICCAT completed a preliminary stock assessment (ICCAT, 2005). Although results suggested that the North Atlantic stock were above biomass in support of MSY, the assessment remained conditional on the assumptions made. These assumptions included (i) estimates of historical shark catch, (ii) the relationship between catch rates and abundance, (iii) the initial state of the stock in 1971, and (iv) various life-history parameters. It was pointed out that the data used for the assessment did not meet the requirements for proper assessment (ICCAT, 2006), and further research and better-resolved data collection was highly recommended.

In 2008, three models were used in stock assessment conducted by ICCAT (ICCAT, 2008 and references cited therein): a Bayesian surplus production model, an age-structured model that did not require catch data (catch-free model), and an age-structured production model. Results with the Bayesian surplus production model produced estimates of stock size well above MSY levels (1.5–2* B_{MSY}), and estimated F to be very low (at F_{MSY} or well below it). The carrying capacity of the stock was estimated so high that the increasing estimated catches (25–62 000 t over the time-series) generated very low F estimates. Sensitivity analyses showed that the stock size estimate was dependent on the weighting assigned to the Irish CPUE series. Equal weighting of this and the other series produced a stock size at around B_{MSY}. Other sensitivity analyses indicated similar results to the base case run, with the stock well above MSY levels.

The age-structured biomass model displayed different results with either a strong decrease in biomass throughout the series to about 30% of virgin levels, or a less pronounced decline. The prior for the virgin biomass assigned high values to a very small number of biomass values but also indicated that the range of plausible values of this parameter has a heavy tail. This is probably because there is not enough information in the data to update the model and thus provide a narrower range of plausible values and thus provide a more precise estimate of the biomass of the stock.

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The age-structured model not requiring catch information estimated that F was higher than F_{MSY} , but still low and that the current SSB estimated at around 83% of virgin levels.

As a consequence of the results in 2008, ICCAT concluded that biomass was estimated to be above the level that would support MSY (ICCAT, 2008). These results agreed with earlier work (ICCAT, 2005). Stock status appeared to be close to unfished biomass levels and fishing mortality rates were well below those corresponding to the level at which MSY is reached. However, IC-CAT (2008) pointed out that the results were heavily dependent on the underlying assumptions. In particular, the choice of catch data to be used, the weighting of CPUE series and various life-history parameters used as input in the model. ICCAT was unable to conduct sensitivity analyses of the input data and assumptions (ICCAT, 2008).

Owing to those weaknesses, no firm conclusions were drawn from the preliminary assessments conducted by ICCAT. ICCAT, 2008 stated that most models used predicted that this stock was not overfished but did not use these results to infer stock status and to provide management advice.

8.9 Stock assessment

The North Atlantic Blue shark stock was assessed by ICCAT in 2015 using two different approaches (see ICCAT, 2015 for more details): Bayesian Surplus Production Model (BSPM) and length-based age-structured models - Stock Synthesis (SS3).

The Bayesian Surplus Production Models adjusted consistently estimated a posterior distribution for *r* that was similar to the prior, and a posterior for *K* with a long right tail with high mean and CV (ICCAT, 2015). The estimated biomass trajectory stayed close to *K* for most runs, and the harvest rate estimate was low (Figure 8.8). The inclusion of a process error in the model did not improve the results. When each CPUE index was fitted separately, the posterior mean of *K* varied and the CVs were large, implying that none of the indices were particularly informative about the value of *K*.

Several SS3 runs were undertaken. Run 4 and 6 (see details below) which utilized multiplication factors to reduce the input sample size assigned to length composition data in the model likelihood resulted in reasonable convergence diagnostics (described below).

Model Run			Model Adjust	ments	
Preliminary Run 1	Natural wei	ghts used in mode	l likelihood		
	Length com	position input san	ple size (n = obse	rved)	
	Abundance	indices (inverse C	V weighting; SCF	(S/2015/151)	
Preliminary Run 2	Same as Pre	liminary Run 1 +	Adjust CV of S9 (ESP-LL-N)	
CV adjustment	Constant CV	of 20% applied t	to S9 (ESP-LL-N)		
Preliminary Run 3	Same as Pre	liminary Run 2 +	Adjust input samp	le size for length	comp
Sample size adjustments	Maximum le	ength composition	input sample size	(n=200)	
Preliminary Run 4	Same as Pre	liminary Run 2 +	Apply variance ad	justment to length	comp.
Fleet	F1	F2	F3	F4	F5
Variance adjustments	0.01	0.01	0.1	0.1	0.1
Preliminary Run 5	Same as Pre	liminary Run 2 +	Apply variance ad	justment to length	comp.
Fleet	F1	F2	F3	F4	F5
Variance adjustments	0.0184	0.0478	0.0261	0.1373	0.2236
Preliminary Run 6	Same as Pre	liminary Run 2 +	Apply variance ad	justment to length	comp.
Fleet	F1	F2	F3	F4	F5
Variance adjustments	0.0019	0.0047	0.0046	0.0573	0.0403

Model fits to CPUE and length composition data were similar for both runs. The fitting to abundance tracked trends well and were within most annual 95% confidence intervals for many abundance indices, including S3 (JPLL-N-e), S4 (JPLL-N-l), S6 (US-Obs-cru), S7 (POR-LL), and S9 (ESP-LL-N) (Figures 8.9–8.10). Model fits tracked trends reasonably well for abundance index S2 (US-Obs), but were often outside annual 95% confidence intervals. Predicted abundance was flat for abundance indices S8 (VEN-LL) and S10 (CTP-LL-N), probably because of large 95% confidence intervals for S8 and high inter-annual fluctuations in the early years for S10. Indices S1 (US-Log) and S5 (IRL-Rec) were only included in the model for exploratory purposes, were not fit in the model likelihood (lambda = 0), and had no influence on model results or predicted values. Model fits to length composition were reasonable for aggregate data (Figure 8.11).

Both run 4 and run 6 resulted in sustainable spawning stock size and fishing mortality rates relative to maximum sustainable yield (Figures 8.12–8.14). However, run 6 (the model run with relatively less weight applied to the length composition data in the model likelihood) resulted in a relatively more depleted stock size, compared to run 4.

Both models suggested sustainable spawning stock size and fishing mortality rates relative to maximum sustainable yield. The model with a relatively lower sample size assigned to the length composition data resulted in a relatively more depleted stock size. However, model fits to length composition were insufficient for annual length composition data, for which a bimodal pattern was evident. This is related to spatial segregation of the population. It was suggested that more work should be done to improve the fits to length composition data before using the model to provide management advice.

8.10 Quality of assessments

At the 2015 ICCAT assessment meeting, considerable progress was made on the integration of new data sources (in particular size data) and modelling approaches (in particular model structure). Uncertainty in data inputs and model configuration was explored through sensitivity analyses, which revealed that results were sensitive to structural assumptions of the models. The production models showed a poor fit to the flat or increasing trends in the CPUE series combined with increasing catches. Overall, assessment results are uncertain (e.g. level of absolute abundance varied by an order of magnitude between models with different structures) and should be interpreted with caution.

For the North Atlantic stock, scenarios with the BSPM estimated that the stock was not overfished (B₂₀₁₃/B_{MSY} = 1.50–1.96) and that overfishing was not occurring (F₂₀₁₃/F_{MSY} = 0.04–0.50). Estimates obtained with SS3 varied more widely, but still predicted that the stock was not overfished (SSF₂₀₁₃/SSF_{MSY} = 1.35–3.45) and that overfishing was not occurring (F₂₀₁₃/F_{MSY} = 0.15–0.75). Comparison of results obtained in the assessment conducted in 2008 and the current assessment revealed that, despite significant differences between inputs and models used, stock status results did not change drastically (B₂₀₀₇/B_{MSY} = 1.87–2.74 and F₂₀₀₇/F_{MSY} = 0.13–0.17 for the 2008 base runs using the BSP and a catch-free age-structured production model).

8.11 Reference points

ICCAT uses F/F_{MSY} and B/B_{MSY} as reference points for stock status of this stock. These reference points are relative metrics rather than absolute values. The absolute values of B_{MSY} and F_{MSY} depend on model assumptions and results and are not presented by ICCAT for advisory purposes.

8.12 Conservation considerations

The global IUCN listing for blue shark is Near Threatened (Rigby *et al.*, 2019), and it has the same listing in European waters, although is listed as Critically Endangered in the Mediterranean Sea (https://www.iucnredlist.org).

Blue shark was listed on Appendix II of the Convention on the Conservation of Migratory Species of Wild Animals (CMS) in 2017. However, it was not subsequently listed on Annex 1 of the Sharks-MoU.

8.13 Management considerations

Based on the scenarios and models explored, ICCAT considered the status of the North Atlantic stock as unlikely to be overfished nor subject to overfishing. However, due to the level of uncertainty, no specific management recommendations have been developed until 2017. Since 2017 Recommendation 16–12 is in place which states: "If the average total catch of the North Atlantic blue shark in any consecutive two years from 2017 onward exceeds the average level observed during the period 2011–2015 (i.e. 39 102 t), the Commission shall review the implementation and effectiveness of these measures. Based on the review and the results of the next stock assessment scheduled for 2021 or at an earlier stage if enough information is provided to SCRS, the Commission shall consider introduction of additional measures".

A further update in 2019 in Recommendation 19-07 is a follows: "If in any year the total catches of the North Atlantic blue shark exceed the TAC, the Commission shall review the implementation of these measures. Based on the review and the results of the next stock assessment scheduled for 2021 or at an earlier stage if enough information is provided to the SCRS, the Commission shall consider introduction of additional measures." In this same Recommendation (19-07) catch limits for the EU, Japan and Morocco were set.

Catch data are highly unreliable. Some CPUE series exist, and where data are available, show a relatively flat trend throughout the time-series, but with high variance. Further work is required to explain the trends and to better quantify removals from the stock.

Catch data are considered incomplete, and underestimated. There have been unaccounted discards and a substantial occurrence of finning over parts of the time series. Data reported to ICES, ICCAT and FAO can vary.

For accurate stock assessments of pelagic sharks, better fishery data are required. In addition, reporting procedures must be strengthened so that all landings are reported, and that landings are reported to species level, rather than generic "shark nei" categories. In the absence of reliable landings and catch data, catch ratios and market information derived from observers can provide useful information for understanding blue shark fishery dynamics.

For the North Atlantic stock, smaller sized blue sharks have been observed to dominate north of 30°N, while larger sized blue sharks dominated south of 30°N. In order to be able to account for the differences in size composition of fish in different areas, future implementations of SS3 should consider this spatial structure in the fleets. This will require estimating fleet and area specific CPUE indices, catch and size distributions. Ideally the model could also be separated by sex.

Blue shark is considered to be one of the most productive sharks in the North Atlantic. As such, it can be expected to be more resilient to fishing pressure than other pelagic sharks. However, the high degree of susceptibility to longline fishing and the poor quality of the information available to assess the stock is a cause for concern. Given the uncertainty of the results and that this

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species is a significant bycatch, especially in tuna and billfish fisheries, there is a need for continued monitoring of the fisheries by observer and port sampling programmes. There are currently no fishery-independent data available for that part of the stock in the ICES area.

8.14 References

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Stock	Country	1978	1979	1981	1982	1983	1984	1985	1986	1987	1988	1989	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999
North Atlantic	Belize																					
	Brazil																					
	Canada								320	147	968	978	680	774	1277	1702	1260	1494	528	831	612	547
	Cape Verde																+					
	China P.R.																					
	Chinese Tai- pei																					
	EU.Denmark											2	2	1	1		1	2	3	1	1	
	EU.Spain																			24497	22504	21811
	EU.France	4	12		9	8	14	39	50	67	91	79	130	187	276	322	350	266	278	213	163	399
	EU.Ireland																					66
	EU.Nether- lands																					
	EU.Portugal												1387	2257	1583	5726	4669	4722	4843	2630	2440	2227
	EU.UK												1				+	12			1	+
	FR.St Pierre et Miquelon																					
	Japan																1203	1145	618	489	340	357
	Mexico																	+				
	Panama																					9
	Senegal																					
	Trinidad & Tobago																					
				204		605	107	341	1112	1400	776	751	829	1080	399	1816	601	641	987	391	447	317

Table 8.1. Blue shark in the North Atlantic. Landings (t) by country 1978–2018 from ICCAT Task I catch data (accessed June 2019). These are considered underestimates, especially prior to 1997.

Stock	Country	1978	1979	1981	1982	1983	1984	1985	1986	1987	1988	1989	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999
	Korea Rep.																					
	Namibia																					
	South Africa																					
	Uruguay																					
	Venezuela																					
N. Atlantic	Total	4	12	204	9	613	121	380	1482	1614	1835	1810	3028	4299	3536	9566	8084	8285	7258	29053	26510	25741
Mediterra- nean	EU.Cyprus																					
	EU.Spain																			146	59	20
	EU.France																					
	EU.Italy																					
	EU.Malta																1	1	1	+	+	+
	EU.Portugal																				2	
	Japan																5	7	1	1		
Mediterra- nean	Total	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	5.581	8.376	1.768	147.95	60.856	20.445
N.ATL AND	MED TOTAL	4	12	204	9	613	121	380	1482	1614	1835	1810	3028	4299	3536	9566	8090	8293	7260	29201	26571	25761

			2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015	2016
	ATN		28174	21709	20066	22951	21742	22359	23217	26927	30723	35198	37178	38084	36786	37202	39881	39502	4202
	ATS		12444	14043	12682	14967	14438	20642	20493	23487	23097	23459	27799	34926	26347	19545	29292	22172	2393
	MED		45	47	17	11	125	72	178	50	81	185	216	40	42	100	235	85	7
TOTAL			40664	35800	32765	37928	36305	43072	43888	50464	53901	58842	65193	73050	63174	56848	69408	62012	6627
Landings	ATN	Longline	27305	20699	19290	22880	21297	22167	23067	26810	30514	35031	36952	37777	36549	36882	39677	38777	4177
		Other surf.	732	905	708	70	380	126	104	63	80	63	59	100	109	74	205	725	25
	ATS	Longline	12444	14042	12678	14961	14339	20638	20434	23417	22708	23453	27785	34531	25878	19375	27457	21355	2330
		Other surf.	0	1	4	6	99	3	59	10	375	6	14	391	264	0	1835	818	62
	MED	Longline	44	47	17	10	43	71	83	48	81	18	50	40	41	68	190	84	7
		Other surf.	1	1	1	0	81	0	95	2	1	167	165	0	0	32	45	1	
Discards	ATN	Longline	137	105	68	0	63	66	45	53	129	102	167	205	127	246	122	124	8
		Other surf.	0	0	0	0	1	0	0	0	1	1	1	2	1	0		+	
	ATS	Longline	0	0	0	0	0	0	0	60	14	0	0	4	206	169	114	122	13
		Other surf.	0	0	0	0	0	0	0	0	0	0	0	0	0	0		6	
Landings	ATN	Barbados																9	
		Belize	0	0	0	0	0	0	0	0	0	114	461	1039	903	1216		4	
		Brazil	7	0	0	0	0	0	0	0	0	0	0	0	0	0		0	
		Canada	624	1162	836	346	965	1134	977	843	0	0	0	0	1	0		0	
		Cape Verde	0	0	0	0	0	0	0	0	0	0	0	0	0	0		0	
		China PR	0	185	104	148	0	0	0	367	109	88	53	109	98	327		1	2
		Chinese Taipei	165	59	0	171	206	240	588	292	110	73	99	148	94	121	81	220	26
		EU.Denmark	2	1	13	5	1	0	0	0	0	0	0	0	0	0		0	
		EU.Spain	24112	17362	15666	15975	17314	15006	15464	17038	20788	24465	26094	27988	28666	28562	25202	30078	2901
		EU.France	395	207	221	57	106	120	99	167	119	84	122	115	31	216	129	259	35
		EU.Ireland	31	66	11	2	0	0	0	0	0	0	0	1	3	2	1	0	

Table 8.1. Cont. Blue shark in the North Atlantic. Landings (t) by country 1978–2018 from ICCAT Task I catch data (accessed June 2019). These are considered underestimates, especially prior to 1997.

		2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015	201
	EU.Netherlands	0	0	0	0	0	0	0	0	0	1	0	0	0	0		0	
	EU.Portugal	2081	2110	2265	5643	2025	4027	4338	5283	6167	6252	8261	6509	3768	3694	2913	3859	78:
	EU.UKi	12	9	6	4	6	5	3	6	6	96	8	10	8	10	10	12	
	FR.St Pierre et Miquelon	0	0	0	0	0	0	0	0	0	1	0	0	0	0		0	
	Japan	273	350	386	558	1035	1729	1434	1921	2531	2007	1763	1227	2437	1808	2034	4011	42
	Korea Rep.	0	0	0	0	0	0	0	0	0	0	0	537	299	327		0	
	Marocco																873	
	Mexico	0	6	0	0	0	0	0	0	0	0	0	0	0	0		0	
	Panama	0	0	0	0	0	0	254	892	613	1575	0	0	0	289		0	
	Senegal	0	0	456	0	0	0	0	43	134	255	56	0	5	12		13	
	St.Vincent and Grenadines																0	
	Suriname	0	0	0	0	0	0	0	0	0	0	0	0	181	281		0	
	Trinidad and Tobago	0	0	6	3	2	1	1	0	2	8	9	11	11	8		4	
	U.S.A.	291	39	0	0	7	2	2	1	8	4	9	65	56	32		31	
	UK.Bermuda	0	0	0	0	0	0	0	0	0	0	0	0	0	0		0	
	Venezuela	43	47	29	40	10	28	12	19	8	73	75	118	98	52		129	
scards ATN	Candada																0	
	Chinese Taipei	0	0	0	0	0	0	0	0	0	0	0	0	21	14	9	5	
	Korea Rep.	0	0	0	0	0	0	0	0	0	0	0	0	0	0		19	
	U.S.A.	137	106	68	0	65	66	45	54	130	103	167	206	106	231		18	
	UK.Bermuda	0	0	0	0	0	0	0	0	0	0	0	0	0	0		82	

Component	Area	Fleet	2017	2018
Landings	ATN	Longline	38509	32654
-		Other surf.	1033	108
	ATS	Longline	27522	3354
		Other surf.	487	66
	MED	Longline	92	54
		Other surf.	13	
Discards	ATN	Longline	133	11
Discalus		Other surf.	0	
	ATS	Longline	218	9
	AIS	Other surf.		
			5	
Landings	ATN	Barbados	7	
		Belize	201	31
		Brazil	0	
		Canada	0	
		Cape Verde	0	
		China PR	2	
		Chinese Taipei		
		EU.Denmark	0	
		EU.Spain	27316	2168
		EU.France	124	9
		EU.Ireland	0	
		EU.Netherlands		
		EU.Portugal	5664	519
		EU.United Kingdom	11	
		FR.St Pierre et Miquelon	0	
		Japan	4460	411
		Korea Rep.	103	9
		•		164
		Marocco	1475	104
		Mexico	0	
		Panama	0	43
		Senegal	4	
		St.Vincent and Grena-		
		Suriname		
		Trinidad and Tobago	2	
		U.S.A.	24	1
		UK.Bermuda	0	
		Venezuela	104	
Discards		Candada	32	7
		Chinese Taipei	34	3
		Korea Rep.	29	
		U.S.A.	38	1
		UK.Bermuda		
Total	ATN	UNDERHAU	39675	3385
	ATS		28232	3430
	MED		105	5450
	Total		68011	6822

Table 8.2. Blue shark in the North Atlantic. Update for 2018 from SCRS (2019).

Table 8.3. Blue shark in the North Atlantic. ICCAT Task I catch data by area for North Atlantic blue shark (accessed 15 June 2021, based on ICCAT data from 18/12/2020; data relate to landings and dead discards). Areas are Azores (AZOR), Canary Islands (CANA), Cape Verde (CVER), Gulf of Mexico (GOFM), Madeira Islands (MDRA), Northeast Atlantic (NE), North Atlantic (NORT), Northwest Atlantic (NW), Northwest Central Atlantic (NWC) and West Tropical Atlantic (WTRO).

Year	AZOR	CANA	CVER	GOFM	MDRA	NE	NORT	NW	NWC	WTRO	Total
1997				1		15363	203	1132	12547	40	29285
1998						12410	246	1047	13051	10	26764
1999				1		11960	617	892	12654	48	26172
2000				1		11857	438	989	14844	44	28174
2001				9		10351	594	687	9440	47	21128
2002				0		9847	490	858	8835	35	20066
2003	639					9617	1208	359	11143	40	23006
2004	526					6423	1515	1020	12245	12	21741
2005	1485		205	0	8	6378	1969	1160	11125	28	22359
2006	1203		186		28	6459	2022	1376	11931	12	23218
2007	1305		259	0	12	6588	4189	1745	12810	20	26927
2008	981		374	0	33	8794	5219	784	14530	10	30725
2009	2051	1	544	0	11	13363		2058	17056	114	35199
2010	3221	47	971	0	34	12870		938	19029	128	37239
2011	1287	4	774	0	18	10749		2132	22939	189	38092
2012	293		281	1		13443		1863	20707	195	36783
2013	110			0	0	13816		1958	20850	353	37087
2014	26			1		14159		1052	21193	149	36579
2015	38			1		18436		318	20680	155	39627
2016	40			0	0	26260		9269	8366	131	44068
2017	15			3	0	22308		9876	7445	17	39664
2018	620			4		17993		8795	6576	6	33995
2019				0	0	15994		5872	5400	12	27279

Table 8.4. Blue shark in the North Atlantic. Length–weight relationships for blue shark from different populations. Lengths in cm, and weights in kg unless specified in equation. W_R = round weight; W_D = dressed weight.

L (cm) W (kg) relationship	Sex	n	Length range (cm)	Source
W _D = (8.04021 x 10–7) L _F ^ 3.23189	С	354	75–250 (L _F)	García-Cortés and Mejuto, 2002
W _R = (3.1841 x 10–6) L _F ^ 3.1313	С	4529		Castro, 1983
W _R = (3.92 x 10–6) LT ^ 3.41	Male	17		Stevens, 1975
W _R = (3.184 x 10–7) L _T ^ 3.20	Female	450		Stevens, 1975
W _R = (3.2 x 10–6) L _F ^ 3.128	С	720		Campana <i>et al.,</i> 2005
W _D = (1.7 x 10–6) L _F ^ 3.205	С	382		Campana <i>et al.,</i> 2005

Table 8.5(a). Blue shark in the North Atlantic. Length–length relationships for male, female blue shark and both sexes combined from the NE Atlantic and Straits of Gibraltar (Buencuerpo *et al.*, 1998). L_s = standard length; L_F = fork length; L_T = total length; L_{UC} = upper caudal lobe length.

Females	Males	Combined
L _F = 1.076 L _S + 1.862 (n = 1043)	L _F = 1.080 L _S + 1.552 (n = 1276)	L _F = 1.079 L _S + 1.668 (n = 2319)
L _T = 1.249 L _S + 7.476 (n = 1043)	L _T = 1.272 L _S + 4.466 (n = 1272)	L _T = 1.262 L _S + 5.746 (n = 2315)
L _{UC} = 0.219 L _S + 4.861 (n = 1038)	L _{UC} = 0.316 L _s + 2.191 (n = 1264)	L _{UC} = 0.306 L _s + 3.288 (n = 2302)
L _T = 1.158 L _F + 5.678 (n = 1043)	L _T = 1.117 L _F + 2.958 (n = 1272)	L _T = 1.167 L _F + 4.133 (n = 2315)

Table 8.5(b). Blue shark in the North Atlantic. Length–length relationships for both sexes combined of blue shark from
various populations and sources.

Stock	Relationship	n	Source
NW Atlantic	L _F = (0.8313) L _T + 1.3908	572	Kohler <i>et al.,</i> 1995
NE Atlantic	$L_F = 0.8203 L_T - 1.061$		Castro and Mejuto, 1995
NW Atlantic	L _F = -1.2 +0.842 L _T	792	Campana et al., 2005
NW Atlantic	L _T = 3.8 + 1.17 L _F	792	Campana et al., 2005
NW Atlantic	$L_{CF} = 2.1 + 1.0 L_{SF}$	782	Campana et al., 2005
NW Atlantic	$L_{SF} = -0.8 + 0.98 L_{CF}$	782	Campana et al., 2005
NW Atlantic	L _F = 23.4 + 3.50 L _{ID}	894	Campana et al., 2005
NW Atlantic	$L_{ID} = -4.3 + 0.273 L_F$	894	Campana et al., 2005

Year	Usobs	JPLLe	North Atla JPLLI	USOLD	PORLL	VENLL	ESPLL	CHTPLI
1957	03003	VELLO	UFLLI	0.98	PONEL	VENEL	LOFLL	CITIPL
1958				0.48				
1959				1.11				
1960				1.18				
1961				1.13				
1962				1.5				
1963				0.7				
1964				0.87				
1965				1.55				
1966				1.27				
1967				1.43				
1968				1.31				
1969				1.96				
1970				0.97				
1971		0.87		1.08				
1972		1.46		1.93				
1973		1.12						
1974		2.62						
1975		1.85		0.88				
1976		1.07		0.75				
1977		1.89		1.82				
1978		1.58		1.06				
1979		1.3		0.860				
1980		2.21		0.830				
1981		2.19		1.050				
1982		2.08		0.780				
1983		1.81		1.010				
1984		1.22		0.680				
1985		1.51		0.740				
1986		1.51		0.480				
1987				0.400				
		2.13						
1988		1.21		0.440				
1989		1.51		0.800				
1990		1.34		0.940				
1991		1.26		1.220				
1992	7.455	1.9		0.63				
1993	11.076	2.43		0.95				
1994	9.717		2.33	0.98		0.047		
1995	10.17		2.1	0.73		0.073		
1996	8.208		2.05	0.47		0.017		
1997	14.439		2.05	1.25	158.14	0.154	156.83	
1998	18.408		1.72	1.16	169.02	0.216	154.45	
1999	6.663		1.89	0.76	149.83	0.117	179.91	
2000	9.541		1.58	0.78	201.44	0.151	213.05	
2000	2.306		1.71	0.70	222.14	0.133	215.63	
2002	2.300		1.37		200.86	0.074	183.94	
2002	1.876		1.97		238.77	0.044	222.88	
2003	9.503		1.97		256.16	0.044	177.27	0.749
2005	3.193		1.9		218.55	0.006	166.82	2.195
2006	4.674		2.16		212.63	0.013	177.11	1.308
2007	9.645		2.18		241.32	0.060	187.06	0.561
2008	8.512		2.48		225.68	0.088	215.80	0.495
2009	8.322		2.46		228.30	0.045	196.08	0.570
2010	13.545		2.45		276.76	0.040	209.03	0.877
2011	21.806		2.37		233.29	0.044	221.13	0.765
2012	8.128		2.6		305.53	0.107	238.00	0.668
2013	7.374		2.09		304.08	0.044	203.49	1.045

Table 8.6. Blue shark in the North Atlantic. Indices of abundance for North and South Atlantic blue shark stocks. Source: ICCAT (2015).

			North Atla		DODU			
Year	Usobs	JPLLe	JPLLI	USOLD	PORLL	VENLL	ESPLL	CHTPLL
1957				0.17				
1958				0.16				
1959				0.25				
1960				0.38				
1961				0.35				
1962				0.27				
1963				0.25				
1964				0.17				
1965				0.17				
1966				0.23				
1967				0.21				
1968				0.21				
1969				0.22				
1970				0.32				
1971		0.53		0.23				
1972		0.39		0.21				
1973		0.45						
1974		0.43						
1975		0.34		0.19				
1976		0.47		0.19				
1976		0.47		0.29				
1978		0.32		0.2				
		0.32		0.11				
1979								
1980		0.29		0.09				
1981		0.36		0.09				
1982		0.36		0.09				
1983		0.37		0.1				
1984		0.50		0.1				
1985		0.44		0.1				
1986		0.39		0.09				
1987		0.35		0.1				
1988		0.49		0.12				
1989		0.44		0.39				
1990		0.49		0.17				
1991		0.47		0.11				
1992	0.31	0.43		0.1				
1993	0.29	0.40		0.09				
1994	0.29		0.50	0.1		1.08		
1995	0.29		0.55	0.1		0.87		
1996	0.50		0.51	0.3		1.90		
1997	0.33		0.52	0.13	0.084	1.50	0.008	
1998	0.35		0.52	0.15	0.004	0.67	0.008	
1999	0.33		0.33	0.13	0.078	0.84	0.008	
2000	0.34		0.49	0.13		0.04	0.008	
2000	0.32		0.28	0.12	0.083	0.74	0.008	
2002	0.39		0.62		0.086	1.03	0.008	
2003	0.37		0.59		0.082	1.26	0.009	0.40
2004	0.30		0.69		0.084	1.53	0.009	0.12
2005	0.35		0.71		0.087	3.88	0.010	0.19
2006	0.31		0.69		0.084	2.24	0.010	0.06
2007	0.32		0.61		0.085	1.35	0.011	0.22
2008	0.32		0.69		0.085	1.16	0.011	0.28
2009	0.31		0.64		0.086	1.56	0.012	0.17
2010	0.31		0.64		0.089	1.54	0.010	0.10
2011	0.29		0.51		0.079	1.51	0.010	0.12
2012	0.34		0.51		0.081	1.00	0.010	0.11
2013	0.31		0.21		0.085	1.84	0.011	0.14

 Table 8.7. Blue shark in the North Atlantic. Coefficients of variation (CVs) for North and South Atlantic blue shark stocks.

 Source: ICCAT (2015).

Area	L∞	k	t0	Sex	Study
North Atlantic	394	0.133	-0.801	Combined	Aasen, 1966
North Atlantic	423	0,11	-1.035	Combined	Stevens, 1975
NW Atlantic	343	0.16	-0.89	Males	Skomal, 1990
NW Atlantic	375	0.15	-0.87	Females	Skomal, 1990
NE Atlantic	377	0.12	-1.33	Combined	Henderson <i>et al.,</i> 2001
North Atlantic	282	0.18	-1.35	Males	Skomal and Natanson, 2002
North Atlantic	310	0.13	-177	Females	Skomal and Natanson, 2002
North Atlantic	287	0.17	-1.43	Combined	Skomal and Natanson, 2003
NW Atlantic	300	0.68	-0.25	Combined	MacNeil and Campana, 2002 (whole ages)
NW Atlantic	302	0.58	-0.24	Combined	MacNeil and Campana, 2002 (section ages)

Table 8.8. Blue shark in the North Atlantic. Von Bertalanffy growth parameters (L^{∞} in cm (L_T), k in years⁻¹, t0 in years) from published studies.

Table 8.9. Blue shark in the North Atlantic. Biological parameters for blue shark.

Parameter	Values	Sample Size	Area	Reference
Reproduction	Placental viviparity			various
Litter size	25–50 (30 average)			various
Size-at-birth	30–50 cm LT			Various
Sex ratio (males: females)	1.5:1		NE Atlantic	García-Cortés and Mejuto, 2002
	1:1.44		NE Atlantic	Henderson <i>et al.,</i> 2001
	1.33:1		NW Atlantic	Kohler <i>et al.,</i> 2002
	1:2.13		NE Atlantic	Kohler <i>et al.,</i> 2002
	1:1.07	801	NE Atlantic (N. coast Spain)	Mejuto and García- Cortés, 2005
	1:0.9	158	NE Atlantic (S. coast Spain)	_
	1:0.38	2187	N central Atlantic	_
	1:0.53	4550	NW Atlantic	_
Gestation period	9–12 months			Campana <i>et al.,</i> 2002
% of females revealing fe- cundation signs	0.74	415	NE Atlantic (N. coast Spain)	Mejuto and García- Cortés, 2005
	0	76	NE Atlantic (S. coast Spain)	
	36.27	601	N central Atlantic	_
	18.15	1573	NW Atlantic	_
% of pregnant females	0	415	NE Atlantic (N. coast Spain)	Mejuto and García- Cortés, 2005
	0	76	NE Atlantic (S. coast Spain)	_
	14.6	601	N central Atlantic	
	9.8	1573	NW Atlantic	_

I

Parameter	Values	Sample Size	Area	Reference
Male age-at-maturity (years)	4–6			various
Female age-at-maturity (years)	5–7			various
Male length-at-maturity	180–280 cm (LF)		NW Atlantic	Campana <i>et al.,</i> 2002
	190–195 cm (LF)			Francis and Duffy, 2005
	201 cm (LF; 50% maturity)		NW Atlantic	Campana <i>et al.,</i> 2005
Female length-at-maturity	220–320 cm (LF)			Campana <i>et al.,</i> 2002
	170–190 cm (LF)			Francis and Duffy, 2005
	> 185 cm (LF)			Pratt, 1979
Longevity (years)	16–20			Skomal and Natanson, 2003
Natural mortality (M)	0.23		Worldwide	Campana <i>et al.,</i> 2005 (mean of various stud- ies)
Productivity (R2m) estimate: intrinsic rebound	0.061 (assuming no fecundity in- crease)		Pacific	Smith <i>et al.,</i> 1998
Potential rate of increase per year	43% (unfished)		NW Atlantic	Campana <i>et al.,</i> 2005
Population doubling time TD (years)	11.4 (assuming no fecundity in- crease)		Pacific	Smith <i>et al.,</i> 1998
Trophic level	4.1	14		Cortés, 1999

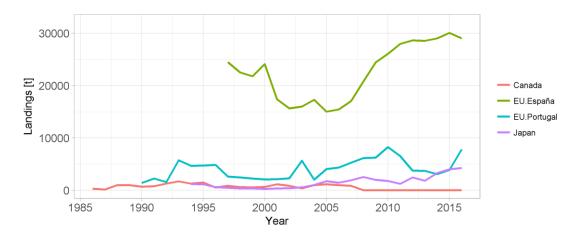


Figure 8.1. Blue shark in the North Atlantic. Preliminary estimates of landings of blue shark in the Atlantic for the four main countries (Source: ICCAT Task I data, Accessed June 2018).

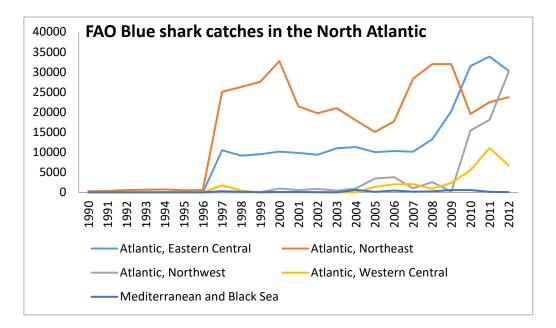


Figure 8.2. Blue shark in the North Atlantic. Preliminary estimates of landings of blue shark in the Atlantic Ocean for the different areas (Source: FAO, 2014).

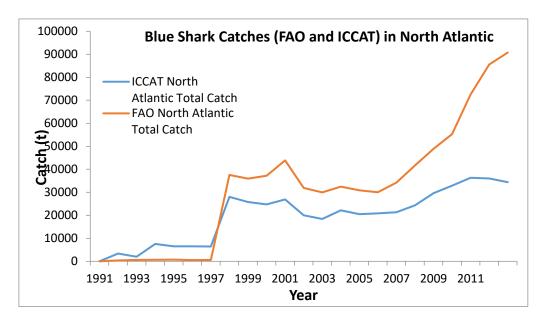
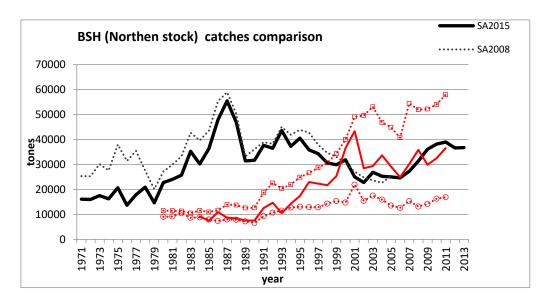


Figure 8.3. Blue shark in the North Atlantic. Blue shark catches in the North Atlantic from FAO and ICCAT data (1990–2013) illustrating the difference between data sources.



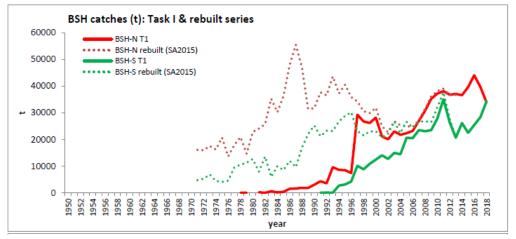


Figure 8.4. Top. Blue shark in the North Atlantic. Comparison of various catch series for the North Atlantic stock of blue shark (1971–2013). In black, the stock assessment catches from the 2008 stock assessment (dotted line) and 2015 estimations (solid line). In red, three catch series obtained using shark-fin ratios with three different approaches (area, effort, target level). Bottom: Update of catches reported to ICCAT (Task I) and estimated by SCRS (SCRS, 2019). Dotted lines are values from the 2008 assessment, solid line those of the 2015 estimates.

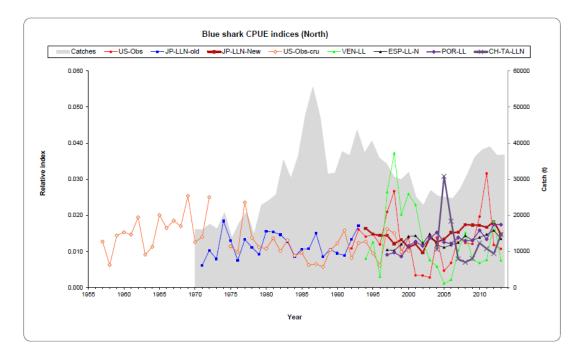


Figure 8.5. Blue shark in the North Atlantic. Indices of abundance and catches. Source: ICCAT (2019).

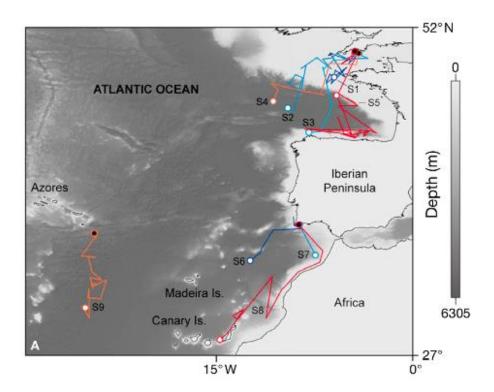


Figure 8.6. Blue shark in the North Atlantic. Pop-off satellite-tagged blue shark movement patterns. (A) General movements overlaid on bathymetry; black circles denote tagging locations and white circles the pop-up/capture locations. (B to J) Individual tracks overlaid on sea surface temperature maps; white circles are geolocated positions with date. Source: Queiroz *et al.* (2010).

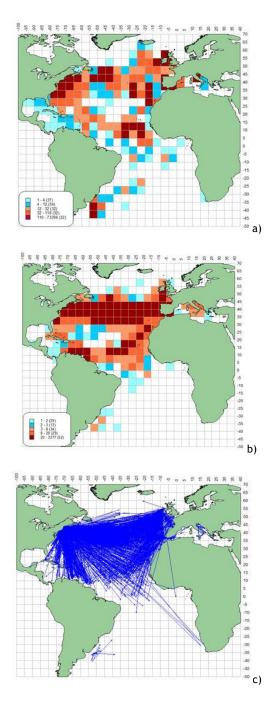


Figure 8.7. Blue shark in the North Atlantic. Blue shark tagging maps, presented by ICCAT (2012), showing (a) density of releases, (b) density of recoveries, and (c) straight line displacement between release and recovery locations.

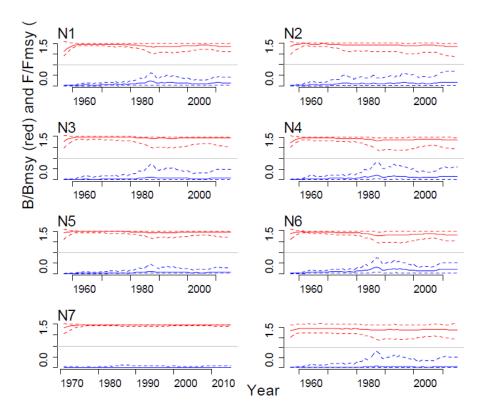


Figure 8.8. Blue shark in the North Atlantic. Estimated biomass relative to B_{MSY} (in red) and harvest rate relative to the MSY level (blue), for the BSP runs. Source: ICCAT (2015).

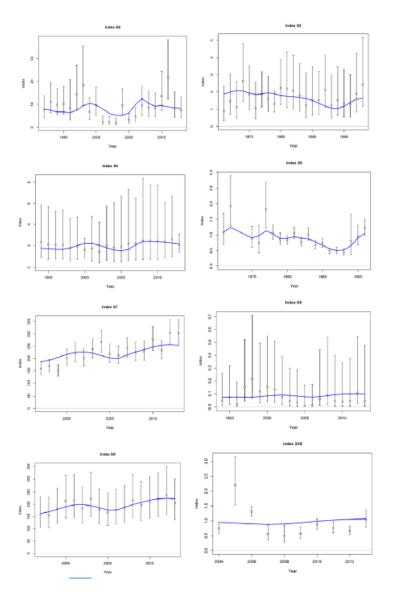


Figure 8.9. Blue shark in the North Atlantic. Preliminary Run 4 observed CPUE (open circles ± 95% confidence intervals assuming lognormal error) and model predicted CPUE (blue line) for abundance indices fit in the model likelihood: S2 (US-Obs, upper left), S3 (JPLL-N-e, upper right), S4 (JPLL-N-l, middle left), S6 (US-Obs-cru, middle right), S7 (POR-LL, middle left), S8 (VEN-LL, middle right), S9 (ESP-LL-N, lower left), and S10 (CTP-LL-N, lower right). Source: ICCAT (2015).

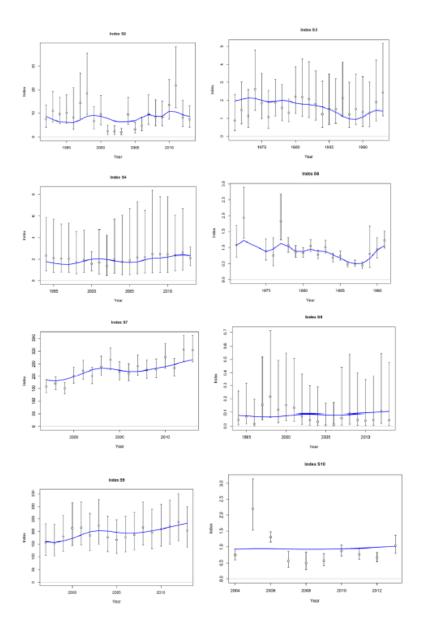
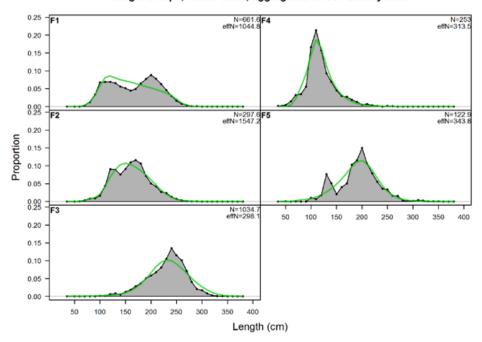


Figure 8.10. Blue shark in the North Atlantic. Preliminary Run 6 observed CPUE (open circles ± 95% confidence intervals assuming lognormal error) and model predicted CPUE (blue line) for abundance indices fit in the model likelihood: S2 (US-Obs, upper left), S3 (JPLL-N-e, upper right), S4 (JPLL-N-l, middle left), S6 (US-Obs-cru, middle right), S7 (POR-LL, middle left), S8 (VEN-LL, middle right), S9 (ESP-LL-N, lower left), and S10 (CTP-LL-N, lower right). Source: ICCAT (2015).



length comps, whole catch, aggregated across time by fleet



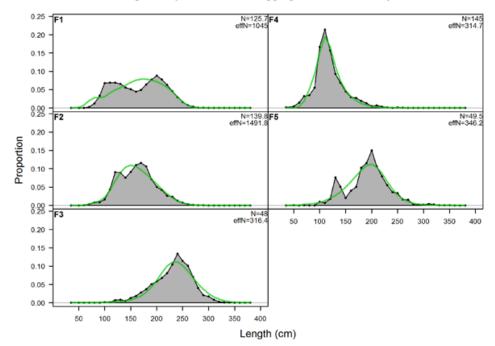


Figure 8.11. Blue shark in the North Atlantic. Model predicted (line) and observed (shaded) aggregated annual length compositions (female + male) for Preliminary Run 4 (upper panel) and Preliminary Run 6 (lower panel). Source: ICCAT (2015).

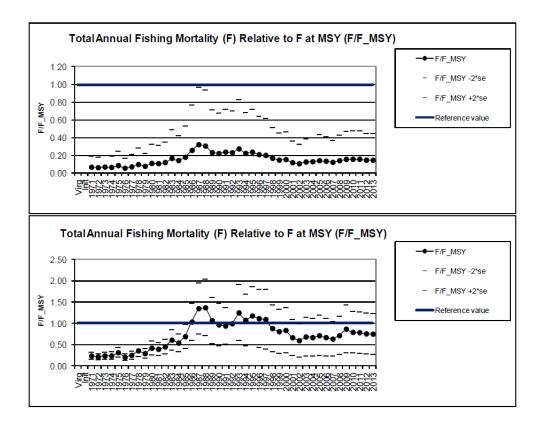


Figure 8.12. Blue shark in the North Atlantic. Estimated annual total exploitation rate in numbers (total fishing mortality for all fleets combined) relative to fishing mortality at MSY (F/F_{MSY}), obtained from Stock Synthesis output for Preliminary Run 4 (upper panel) and Preliminary Run 6 (lower panel). Source: ICCAT (2015).

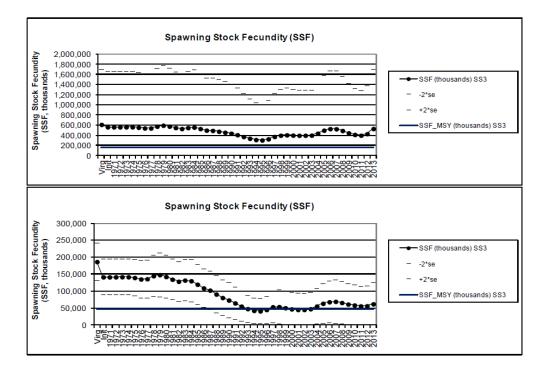


Figure 8.13. Blue shark in the North Atlantic. Estimated spawning stock size (spawning stock fecundity, SSF) along with approximate 95% asymptotic standard errors (± 2 *s.e.) relative to spawning stock size at MSY (SSFMSY) for Preliminary Run 4 (upper panel) and Preliminary Run 6 (lower panel). Source: ICCAT (2015).

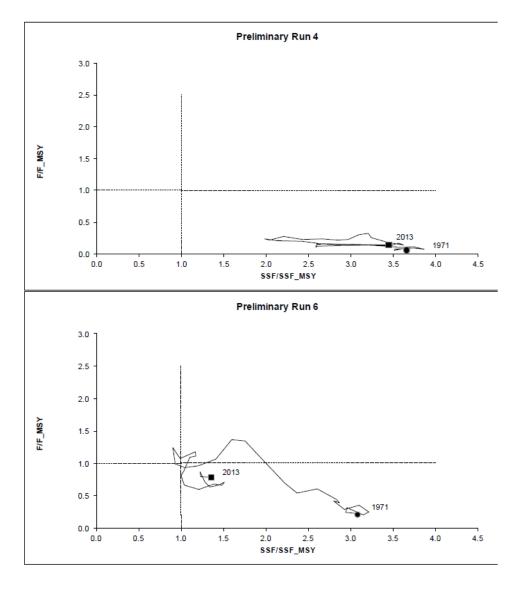


Figure 8.14. Blue shark in the North Atlantic. Kobe Phase plots for Preliminary Run 4 (upper panel) and Preliminary Run 6 (lower panel). The circle indicates the position of the start year of the model (1971) and the square represents the end year of the model (2013). The horizontal (dotted) line identifies the fishing mortality reference at maximum sustainable yield (F_{MSY}). The vertical (dotted) line identifies the reference spawning stock fecundity at maximum sustainable yield (SSF_{MSY}). Source: ICCAT (2015).