## 8 Northeast Atlantic Mackerel

### 8.1 ICES Advice and International Management Applicable to 2020

From 2001 to 2007, the internationally agreed TACs covered most of the distribution area of the Northeast Atlantic mackerel. From 2008 to 2014, no agreement was reached among the Coastal States on the sharing of the mackerel quotas. In 2014, three of the Coastal States (European Union, Norway and the Faroe Islands) agreed on a Management Strategy for 2014 to 2018. In November 2018, the agreement from 2014 was extended for two further years until 2020. No agreement on the share of the stock has been reached after Brexit for 2021. Despite various agreements, the total declared quotas in each of the years 2015 to 2020 all exceeded the TAC advised by ICES. An overview of the declared quotas and transfers for 2021, as available to WGWIDE, is given in the text table below. Total removals of mackerel are expected to be approximately 1.2 million tonnes in 2021, exceeding the ICES advice for 2021 by about 347000 t ( $41 \%$ ).

| Estimation of 2021 catch | Tonnes | Reference |
| :---: | :---: | :---: |
| EU quota | 200179 | NEAFC HOD 21/22 |
| UK quota | 222288 | Department for Environment Food \& Rural Affairs (UK). April 2021 |
| Norwegian quota | 298299 | NEAFC HOD 21/22 |
| Inter-annual quota transfer 2020->2021 (NO) | -10 210 | NEAFC HOD 21/22 |
| Russian quota | 120423 | NEAFC HOD 21/22 |
| Discards | 9280 | Previous years estimate |
| Icelandic expected catch | 120000 | WGWIDE |
| Faroese quota | 167048 | Faroese Fisheries Ministry regulations No. 85 and 115/2021 |
| Inter-annual quota transfer 2020->2021 (FO) | 33796 | Faroese Fisheries Ministry regulations No. 85 and 115/2021 |
| Greenland expected catch | 38000 | Ministry of Fisheries, Hunting and Agriculture in Greenland |
| Total expected catch (incl. discards) ${ }^{1,2}$ | 1199103 |  |

${ }^{1}$ No estimates of banking from 2020 to 2021.
${ }^{2}$ Quotas refer to claims by each party for 2021 and include exchange to other parties
The quota figures and transfers in the text table above were based on various national regulations, official press releases, and discard estimates.

Various international and national measures to protect mackerel are in operation throughout the mackerel catching countries. Refer to Table 8.2.4.1 for an overview.

### 8.2 The Fishery

### 8.2.1 Fleet Composition in 2020

The total fleet can be considered to consist of the following components:
Freezer trawlers. These are commonly large vessels (up to 150 m ) that usually operate a single mid-water pelagic trawl, although smaller vessels may also work as pair trawlers. These vessels are at sea for several weeks and sort and process the catch on board, storing the mackerel in frozen 20 kg blocks. The Dutch, German and the majority of the French and English fleets consist of these vessels which are owned and operated by a small number of Dutch companies. They fish in the North Sea, west of the UK and Ireland and also in the English Channel and further south along the western coast of France. Russian freezer trawlers fish for mackerel during the summer (June-September) in the Norwegian Sea in Division 2.a, mainly inside the NEAFC regulatory area. Part of the Icelandic fishery is in Division 5.a and in some years in 14.b.

Purse seiners. The majority of the Norwegian catch is taken by these vessels, targeting mackerel overwintering close to the Norwegian coastline. The largest vessels ( $>20 \mathrm{~m}$ ) used refrigerated seawater (RSW), storing the catch in tanks containing RSW. Smaller purse seiners use ice to chill their catch which they take on prior to departure. A purse seine fleet is also the most important component of the Spanish fleet. They are numerous and target mackerel early in the year close to the northern Spanish coast. These are dry hold vessels, chilling the catch with ice. Denmark also has a purse seine fleet operating in the northern North Sea.

Pelagic trawlers. These vessels vary in size from $20-100 \mathrm{~m}$ and operate both individually and as pairs. The largest of the pelagic trawlers use RSW tanks for storage. Iceland, Greenland, Faroes, Scotland and Ireland fish mackerel using pelagic trawlers. Scottish and Icelandic vessels mostly operate as single trawlers whereas Ireland and Faroese vessels tend to use pair trawls. Spain also has a significant trawler fleet which target mackerel with a demersal trawl in Subarea 8 and Division 9.a.N.

Lines and jigging. Norway and England have handline fleets operating inshore in the Skagerrak (Norway) and in Divisions 7.e/f (England) around the coast of Cornwall, where other fishing methods are not permitted. Spain also has a large artisanal handline fleet as do France and Portugal. A small proportion of the total catch reported by Scotland (Divisions $4 . a$ and 4.b) and Iceland (Division 5.a) is taken by a handline fleet.

Gillnets. Gillnet fleets are operated by Norway and Spain.

### 8.2.2 Fleet Behaviour in 2020

The northern summer fishery in Subareas 2 and 5 continued in 2020. There was no fishery in Subarea 14. The Russian freezer trawler fleet operates over a wide area in northern international waters. This fleet targets herring and blue whiting in addition to mackerel. In the third quarter of 2020 the Russian vessels took the vast majority of their catch in Division 2.a.

Total catches from Icelandic vessels were similar to those in recent years and were in excess of 100 kt . The majority of the catch was taken in Division 2.a in quarter 3 of 2020, with catch also taken in 5.a in waters to the south, east and west of Iceland. In 2020 Greenland targeted mackerel in Division 2.a with no catch taken from 14.b. In 2019 Greenland fished in 14.b and in 2018 both Greenland and Iceland reported landings from this area. Catches from Greenland have decreased again in 2020 to 27 kt , down from 30 kt in 2019 and almost 63 kt in 2018. The Faroese fleet targeted mackerel during late summer and early autumn with nearly half of the catches taken in 2.a and 4.a. The remaining catch was taken in quarter 1 mainly in $4 . a$ and some in $6 . a$.

Fishing in the North Sea and west of the British Isles followed a traditional pattern, targeting mackerel on their spawning migration from the Norwegian deep in the northern North Sea, westwards around the north coast of Scotland and down the west coast of Scotland and Ireland. The majority of the Irish mackerel fishery took place in quarter 1 along the west coast of Scotland and Ireland, with the Scottish fleet operating in the same area at this time. The Scottish fishery in quarter 4 was more concentrated in the North Sea.

In 2020 the Spanish fishery started at the beginning of March, as in previous years.

### 8.2.3 Recent Changes in Fishing Technology and Fishing Patterns

Northeast Atlantic mackerel, as a widely distributed species, is targeted by a number of different fishing métiers. Most of the fishing patterns of these métiers have remained unchanged during the most recent years, although the timing of the spawning migration and geographical distribution can change from year to year and this affects the fishery in various areas.

The most important changes in recent years are related to the geographical expansion of the northern summer fishery (Subareas 2,5 and 14) and changes in southern waters due to stricter TAC compliance by Spanish authorities. In 2020 the northern summer fishery did not extend as far west as in previous years.

As a result of this expansion, Icelandic vessels have increased effort and catch dramatically in recent years from 4 kt in 2006 to an average 159 kt annually since 2011. This fishery operates over a wide area E, NE, SE, S and SW of Iceland. Since 2011, there has been less fishing activity to the north and north-east and an increase in catches taken south and west of Iceland. Greenland has reported catches from Division 14.6 since 2011, and reached the biggest catch by this fleet to date in 2014, with a catch of 78 kt. In 2020 the catch reported from Greenland was mainly from Division 2.a.

### 8.2.4 Regulations and their Effects

An overview of the major existing technical measures, effort controls and management plans are given in Table 8.2.4.1. Note that there may be additional existing international and national regulations that are not listed here.

Between 2010 and 2020 no overarching Coastal States Agreement/NEAFC Agreement was in place and no overall international regulation on catch limitation was in force. In 2014, three of the Coastal States (The EU, Faroes and Norway) agreed on a Management Strategy for 2015 and the subsequent five years. In November 2018, the agreement from 2014 was extended for two more years until 2020. However, the total declared quotas taken by all parties since 2015 have greatly exceeded the TAC advised by ICES (see Section 8.1). Currently there is no agreement on a management strategy covering all parties fishing mackerel.
Management aimed at a fishing mortality in the range of $0.15-0.20$ in the period 1998-2008. In 2008 the Coastal states agreed a long term management plan which aimed at a fishing mortality in the range $0.20-0.22$. The fishing mortality realised during 1998-2008 was in the range of 0.27 to 0.46 . Implementation of the management plan resulted in a reduced fishing mortality and increased biomass. The last agreed management plan was in 2017 (ICES, 2017a). During the Coastal States' negotiations in 2019 for 2020, it was recognised that the F and B were outdated after the recent MSE on mackerel (ICES, 2019). Therefore, the Coastal States used Fmsy as reference F in setting their TAC for 2020. At the same time, they requested ICES to evaluate a new management plan for mackerel, which was finally evaluated by ICES in 2020. However, the Coastal States have not considered the response from ICES yet. Since 2008 catches have greatly exceeded those given by the plan.

The measures advised by ICES to protect the North Sea spawning component aim at setting the conditions for making a recovery of this component possible. Before the late 1960s, the North Sea spawning biomass of mackerel was estimated at above 2.5 million tonnes. The collapse of mackerel in the North Sea in the late 1960s was most likely driven by very high catches and associated fishing mortality. However, the lack of recovery of mackerel in the North Sea was probably associated with unfavourable environmental conditions, particularly reduced temperatures (unfavourable for spawning), lower zooplankton availability in the North Sea and increased windstress induced turbulence (Jansen, 2014). These unfavourable environmental conditions probably led the mackerel to spawn in western waters instead of in the North Sea.

A review of the mackerel in the North Sea, carried out during WKWIDE 2017 (ICES, 2017) concluded that Northeast Atlantic mackerel should be considered as a single population (stock) with individuals that show stronger or weaker affinity for spawning in certain parts of the spawning area. Management should ensure that fisheries do not decrease genetic and behavioural diversity, since this could reduce future production. Protection of mackerel that tend to spawn in the north-eastern parts of the spawning area is therefore still advisable to some extent.

In the southern area, a Spanish national regulation affecting mackerel catches of Spanish fisheries has been implemented since 2010. In 2015, fishing opportunities were distributed by region and gear and for the bottom trawl fleet, by individual vessel. This year, Spanish mackerel fishing opportunities in Divisions 8.c and 9.a were established at 39674 t resulting from the quota established (Commission Regulation (EU) No 104/2015). This was reduced by 9797 t due to the scheduling payback quota due to overfishing of the mackerel quota allocated to Spain in 2010 (Commission Regulation No 976/2012).

Within the area of the southwest Mackerel Box off Cornwall in southern England only handliners are permitted to target mackerel. This area was set up at a time of high fishing effort in the area in 1981 by Council Regulation to protect juvenile mackerel, as the area is a well-known nursery. The area of the box was extended to its present size in 1989.

Additionally, there are various other national measures in operation in some of the mackerel catching countries.

The first phase of a landing obligation came into force in 2015 for all EU vessels in pelagic and industrial fisheries. Since 2019, all species that are managed through TACs and quotas must be landed under the obligation unless there is a specific exemption such as de minimis. There are de minimis exemptions for mackerel caught in bottom-trawl fisheries in the North Western Waters (EC 2018/2034) and in the North Sea (EC 2018/2035).

### 8.3 Quality and Adequacy of sampling Data from Commercial Fishery

The sampling of the commercial catch of Northeast Atlantic mackerel is summarised below:

| Year | WG Total Catch |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| (t) | \% catch covered <br> by sampling pro- <br> gramme* | No. <br> Samples | No. | Measured | Aged |
| 1992 | 760000 | 85 | 920 | 77000 | 11800 |
| 1993 | 825000 | 83 | 890 | 80411 | 12922 |
| 1994 | 822000 | 80 | 807 | 72541 | 13360 |


| Year | WG Total Catch <br> (t) | \% catch covered by sampling programme* | No. <br> Samples | No. <br> Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 755000 | 85 | 1008 | 102383 | 14481 |
| 1996 | 563600 | 79 | 1492 | 171830 | 14130 |
| 1997 | 569600 | 83 | 1067 | 138845 | 16355 |
| 1998 | 666700 | 80 | 1252 | 130011 | 19371 |
| 1999 | 608928 | 86 | 1109 | 116978 | 17432 |
| 2000 | 667158 | 76 | 1182 | 122769 | 15923 |
| 2001 | 677708 | 83 | 1419 | 142517 | 19824 |
| 2002 | 717882 | 87 | 1450 | 184101 | 26146 |
| 2003 | 617330 | 80 | 1212 | 148501 | 19779 |
| 2004 | 611461 | 79 | 1380 | 177812 | 24173 |
| 2005 | 543486 | 83 | 1229 | 164593 | 20217 |
| 2006 | 472652 | 85 | 1604 | 183767 | 23467 |
| 2007 | 579379 | 87 | 1267 | 139789 | 21791 |
| 2008 | 611063 | 88 | 1234 | 141425 | 24350 |
| 2009 | 734889 | 87 | 1231 | 139867 | 28722 |
| 2010 | 877272 | 91 | 1241 | 124695 | 29462 |
| 2011 | 948963 | 88 | 923 | 97818 | 22817 |
| 2012 | 899551 | 89 | 1216 | 135610 | 38365 |
| 2013 | 938299 | 89 | 1092 | 115870 | 25178 |
| 2014 | 1401788 | 90 | 1506 | 117250 | 43475 |
| 2015 | 1215827 | 88 | 2132 | 137871 | 24283 |
| 2016 | 1100135 | 89 | 2200 | 149216 | 21456 |
| 2017 | 1159641 | 87 | 2183 | 151548 | 24104 |
| 2018 | 1023144 | 83 | 1858 | 139590 | 20703 |
| 2019 | 839727 | 88 | 1835 | 141561 | 17646 |
| 2020 | 1039513 | 87 | 1430 | 142991 | 15685 |

Overall sampling effort in 2020 was similar to previous years with $87 \%$ of the catch sampled. It should be noted that this proportion is based on the total sampled catch. Nations with large, directed fisheries are capable of sampling $100 \%$ of their catch which may conceal deficiencies in sampling elsewhere.

The 2020 sampling levels by country are shown below.

| Country | Official catch | \% WG catch covered by sampling programme | No. <br> Samples | No. <br> Measured | No. <br> Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 124 | $0 \%$ |  |  |  |
| Denmark | 38589 | $90 \%$ | 14 | 1515 | 967 |
| Faroe Islands | 69064 | $98 \%$ | 12 | 726 | 625 |
| France | 21936 | $0 \%$ |  |  |  |
| Germany | 25030 | 65 \% | 88 | 15351 | 716 |
| Greenland | 26577 | $100 \%$ | 42 | 1998 | 88 |
| Iceland | 151534 | $99 \%$ | 112 | 4895 | 2755 |
| Ireland | 74232 | $99 \%$ | 47 | 8937 | 2061 |
| Lithuania | 815 | $0 \%$ |  |  |  |
| Netherlands | 30321 | 62 \% | 35 | 2633 | 861 |
| Norway | 211672 | $96 \%$ | 65 | 2280 | 1776 |
| Poland | 5302 | $0 \%$ |  |  |  |
| Portugal | 4799 | 12 \% | 101 | 2525 | 988 |
| Russia | 128817 | 100 \% | 201 | 64339 | 1349 |
| Spain | 34613 | $99 \%$ | 622 | 30510 | 2223 |
| Sweden | 3672 | $0 \%$ |  |  |  |
| UK (England \& Wales) | 30430 | 1 \% | 54 | 3165 | 227 |
| UK (Northern Ireland) | 14855 | $34 \%$ | 1 | 166 | 49 |
| UK (Scotland) | 167131 | 89 \% | 36 | 3951 | 1000 |

The majority of countries achieved a high level of sampling coverage. Belgian catches consist of by-catch in the demersal fisheries in the North Sea. France supplied a quantity of length-frequency data to the working group which can be utilised to characterise the selection of the fleet but requires an allocation of catch at age proportions from another sampled fleet in order to raise the data for use in the assessment. Sweden, Lithuania and Poland did not supply sampling information in 2020. Portugal sampled landings from 9.a only. England only samples landings from the handline fleet operating off the Cornish coast, representing only a small proportion of the national catch, the remainder reported from freezer trawlers. Cooperation between the Dutch and German sampling programmes (which sampled $65 \%$ and $62 \%$ respectively) is designed to
provide complete coverage for the freezer trawlers operating under these national flags and also those of England and France. Catch sampling levels per ICES Division (for those with a WG catch of $>100 \mathrm{t}$ ) are shown below.

| Division | Official Catch (t) | WG Catch (t) | No. Samples | No. Measured | No Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 11 | 11 | 0 | 0 | 0 |
| 2.a | 310223 | 310223 | 318 | 69611 | 3424 |
| 3.a | 567 | 567 | 0 | 0 | 0 |
| 3.b | 16 | 16 | 0 | 0 | 0 |
| 3.c | 4 | 4 | 0 | 0 | 0 |
| 3.d | 19 | 19 | 0 | 0 | 0 |
| 4.a | 450720 | 450720 | 228 | 26072 | 5480 |
| 4.b | 5024 | 5024 | 0 | 0 | 0 |
| 4.c | 861 | 861 | 0 | 0 | 0 |
| 5.a | 44867 | 44867 | 44 | 1979 | 1074 |
| 5.b | 1879 | 1879 | 0 | 0 | 0 |
| 6.a | 130903 | 130903 | 40 | 6206 | 1355 |
| 6.b | 15 | 15 | 0 | 0 | 0 |
| 7.a | 5 | 5 | 0 | 0 | 0 |
| 7.b | 20281 | 20281 | 15 | 2261 | 622 |
| 7.c | 191 | 191 | 1 | 51 | 25 |
| 7.d | 5637 | 5637 | 0 | 0 | 0 |
| $7 . \mathrm{e}$ | 8652 | 8652 | 55 | 3278 | 252 |
| 7.f | 260 | 260 | 0 | 0 | 0 |
| 7.9 | 37 | 37 | 0 | 0 | 0 |
| 7.h | 7 | 7 | 0 | 0 | 0 |
| 7.j | 13629 | 13629 | 5 | 383 | 135 |
| 7.k | 1 | 1 | 0 | 0 | 0 |
| 8.a | 2688 | 2688 | 0 | 0 | 0 |
| 8.b | 4727 | 4727 | 185 | 5150 | 389 |
| 8.c | 24128 | 24128 | 47 | 428 | 639 |
| 8.c.E | 11328 | 11328 | 316 | 24466 | 704 |


| Division | Official Catch (t) | WG Catch (t) | No. Samples | No. Measured | No Aged |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 8.d | 754 | 754 | 0 | 0 | 0 |
| $9 . \mathrm{a}$ | 2070 | 2070 | 176 | 3106 | 1586 |
| 9.6 | 2 | 2 | 0 | 0 | 0 |
| 12.c | 6 | 6 | 0 | 0 |  |

In general, areas with insufficient sampling have relatively low levels of catch.

### 8.4 Catch Data

### 8.4.1 ICES Catch Estimates

In 2021 the catch data time series was revised due to additional catch data reported from Division 8.c and the removal of logbook discard data from the working group catch. The led to new working group catch figures as well as a revised catch numbers at age and mean weights at age time series from 2010-2019.

The additional catch in Division 8.c was unsampled. Division 8.c was well sampled by other countries and these samples were allocated to the unsampled catch. For most years and ages, the differences between the previous and the revised catch numbers at age is less than $1 \%$. For the years and ages when the difference is higher this is due to the proportions at age in the sampled catch.

The logbook discard data reported in 2018 and 2019 were submitted from countries that also submitted discard data from observer programmes. It is not known if logbook registered discards are consistently recorded because the reporting of this data is not mandatory and there is a possibility of double counting. It was therefore decided to remove the logbook registered discards and only use the estimates from observer programme. Again, the differences in the previous estimates and the revised estimates was very small. The highest difference was for ages 0 and 1 in 2018 and this was because of the proportions at age in the discard samples that were used in allocations.

The total ICES estimated catch for 2020 was 1039513 an increase of 199786 t on the estimated catch in 2019. Catches increased substantially from 2006-2010 and have averaged 1040 kt since 2011.

The combined 2020 TAC, arising from agreements and autonomous quotas, amounts to 1090879 t . The ICES catch estimate ( 1039513 t ) represents an undershoot of this but is still above the ICES advice of 992064 t . The combined fishable TAC for 2021, as best ascertained by the Working Group (see Section 8.1), amounts to 1199103 t .

Catches reported for 2020 and in previous Working Group reports are considered to be best estimates. In most cases, catch information comes from official logbook records. Other sources of information include catch processors. Some countries provide information on discards and slipped catch from observer programs and compliance reports. In several countries discarding is illegal. Spanish data is based on the official data supplied by the Fisheries General Secretary (SGP) but supplemented by scientific estimates which are recorded as unallocated catch in the ICES estimates.

The text table below gives a brief overview of the basis for the ICES catch estimates.

| Country | Official Log Book | Other Sources | Discard Information |
| :---: | :---: | :---: | :---: |
| Denmark | Y (landings) | Y (sale slips) | Y |
| Faroe ${ }^{1}$ | $Y$ (catches) | Y (coast guard) | NA |
| France | Y (landings) |  | Y |
| Germany | Y (landings) |  | Y |
| Greenland | $Y$ (catches) | Y (sale slips) | Y |
| Iceland ${ }^{1}$ | Y (landings) |  | NA |
| Ireland | Y (landings) |  | Y |
| Netherlands | Y (landings) | Y | Y |
| Norway ${ }^{1}$ | Y (catches) |  | NA |
| Portugal |  | Y (sale slips) | Y |
| Russia ${ }^{1}$ | Y (catches) |  | NA |
| Spain | Y | Y | Y |
| Sweden | Y (landings) |  | Y |
| UK | Y (landings) | Y | Y |

${ }^{1}$ For these nations a discarding ban is in place such that official landings are considered to be equal to catches.
The Working Group considers that the estimates of catch are likely to be an underestimate for the following reasons:

- Estimates of discarding or slipping are either not available or incomplete for most countries. Anecdotal evidence suggests that discarding and slipping can occur for a number of reasons including high-grading (larger fish attract a premium price), lack of quota, storage or processing capacity and when mackerel is taken as by-catch.
- Confidential information suggests substantial under-reported landings for which numerical information is not available for most countries. A study carried out in 2010 indicated considerable uncertainty in true catch figures (Simmonds et al., 2010) for the period studied.
- Estimates of the magnitude and precision of unaccounted mortality suggests that, on average for the period prior to 2007 , total catch related removals were equivalent to 1.7 to 3.6 times the reported catch (Simmonds et al., 2010).
- Reliance on logbook data from EU countries implies (even with $100 \%$ compliance) a precision of recorded landings of 89 \% from 2004 and $82 \%$ previous to this (Council Regulation (EC) Nos. 2807/83 \& 2287/2003). Given that over reporting of mackerel landings is unlikely for economic reasons; the WG considers that the reported landings may be an underestimate of up to $18 \%$ ( $11 \%$ from 2004), based on logbook figures. Where inspections were not carried out there is a possibility of a $56 \%$ under reporting, without there being an obvious illegal record in the logsheets. Without information on the percentage of the landings inspected it is not possible for the Working Group to evaluate the underestimate in its figures due to this technicality.
- The accuracy of logbooks from countries outside the EU has not been evaluated by WGWIDE. Monitoring of logbook records is the responsibility of the national control and enforcement agencies.

The total catch as estimated by ICES is shown in Table 8.4.1.1. It is broken down by ICES area group and illustrates the development of the fishery since 1969.

## Discard Estimates

With a few exceptions, estimates of discards have been provided to the Working Group for the ICES Subareas and Divisions 6, 7/8.a,b,d,e and 3/4 (see Table 8.4.1.1) since 1978. Historical discard estimates were revised during the data compilation exercise undertaken for the 2014 benchmark assessment (ICES, 2014). The Working Group considers that the estimates for these areas are incomplete. In 2020, discard data for mackerel were provided by France, Ireland, Spain, Portugal, Denmark, England, Scotland and Sweden. Total discards amounted to 9280 t which is an increase from 2019. Higher discards were reported by UK England and Wales mainly from one fleet. The German, Dutch and Portuguese pelagic discard monitoring programmes did not record any instances of discarding of mackerel. Estimates from the other countries supplying data include results from the sampling of demersal fleets.

Age-disaggregated discard data was limited in 2020 due to reduced sampling opportunities as a result of COVID but data available indicates that, in Division 8.b the majority of discarded fish were aged 0 to 3. In Divisions 8.c and 9.a, the majority of the discarded fish were 0 group.

Discarding of small mackerel has historically been a major problem in the mackerel fishery and was largely responsible for the introduction of the south-west mackerel box. In the years prior to 1994, there was evidence of large-scale discarding and slipping of small mackerel in the fisheries in Division 2.a and Subarea 4, mainly because of the very high prices paid for larger mackerel (>600 g) for the Japanese market. This factor was put forward as a possible reason for the very low abundance of the 1991 year-class in the 1993 catches. Anecdotal evidence from the fleet suggests that since 1994, discarding/slipping has been reduced in these areas.
In some of the horse mackerel directed fisheries, e.g., those in Subareas 6 and 7, mackerel is taken as by-catch. Reports from these fisheries have suggested that discarding may be significant because of the low mackerel quota relative to the high horse mackerel quota, particularly in those fisheries carried out by freezer trawlers in the fourth quarter. The level of discards is greatly influenced by the market price and by quotas.

### 8.4.2 Distribution of Catches

A significant change in the fishery took place between 2007 and 2009 with a greatly expanded northern fishery becoming established. This fishery has continued to the present but with a clear tendency for an eastern retraction, especially from the Greenlandic area and also western parts of the Icelandic area in the most recent three years. Of the total catch in 2020, Norway accounted for the greatest proportion (20 \%) followed by Scotland (16 \%), Iceland (15 \%), Russia (12 \%), Ireland (7\%) and Faroes (7 \%). In the absence of an international agreement, Greenland, Iceland and Russia declared unilateral quotas in 2020. Russia and Iceland both had catches over 100 kt with Faroes catching 69 kt . Greenlandic catches decreased again from 30 kt to 27 kt . Scotland had catch in excess of 100 kt and Ireland caught 74 kt . Denmark had catches of around 35 kt . The Netherlands and Spain caught around 30 and 34 kt , respectively while UK England had increased catches in 2020 to 30 kt . German catch also increased to 25 kt . France had catches of the order of 22 kt .

In 2020, catches in the northern areas (Subareas 1, 2, 5, 14) amounted to 356985 t (see Table 8.4.2.1), an increase of 11966 t on the 2019 catch. Icelandic, Norwegian and Russian catches were
all over 100 kt . Catches from Division 2.a accounted for $30 \%$ of the total catch in 2020, similar to 2019. Almost all the Russian catch in 2020 was taken in Division 2.a. The wide geographical distribution of the fishery noted in previous years has continued.

The time series of catches by country from the North Sea, Skagerrak and Kattegat (Subarea 4, Division 3.a) is given in Table 8.4.2.2. Catches in 2020 amounted to 457211 t and represents a significant increase of $149164 t$ from the 2019 catch figure ( 308047 t ). The majority of the catch is from Subarea 4 with small catches were also reported in Divisions 3.a-d.

Catches in the western area (Subareas 6, 7 and Divisions 8.a,b,d and e) increased in 2020 to 187788 t . This is an increase of around 26000 t from 2019. The catches are detailed in Table 8.4.2.3.

Table 8.4.2.4 details the catches in the southern areas (Divisions 8.c and 9.a) which are taken almost exclusively by Spain and Portugal. The reported catch of 37529 t represents an increase of almost 13000 t from 2019. The catch is above the long-term average.

The distribution of catches by quarter (\%) is described in the text table below:

| Year | Q1 | Q2 | Q3 | Q4 |
| :---: | :---: | :---: | :---: | :---: |
| 1990 | 28 | 6 | 26 | 40 |
| 1991 | 38 | 5 | 25 | 32 |
| 1992 | 34 | 5 | 24 | 37 |
| 1993 | 29 | 7 | 25 | 39 |
| 1994 | 32 | 6 | 28 | 34 |
| 1995 | 37 | 8 | 27 | 28 |
| 1996 | 37 | 8 | 32 | 23 |
| 1997 | 34 | 11 | 33 | 22 |
| 1998 | 38 | 12 | 24 | 27 |
| 1999 | 36 | 9 | 28 | 27 |
| 2000 | 41 | 4 | 21 | 33 |
| 2001 | 40 | 6 | 23 | 30 |
| 2002 | 37 | 5 | 29 | 28 |
| 2003 | 36 | 5 | 22 | 37 |
| 2004 | 37 | 6 | 28 | 29 |
| 2005 | 46 | 6 | 25 | 23 |
| 2006 | 41 | 5 | 18 | 36 |
| 2007 | 34 | 5 | 21 | 40 |
| 2008 | 34 | 4 | 35 | 27 |
| 2009 | 38 | 11 | 31 | 20 |


| Year | Q1 | Q2 | Q3 | Q4 |
| :---: | :---: | :---: | :---: | :---: |
| 2010 | 26 | 5 | 54 | 15 |
| 2011 | 22 | 7 | 54 | 17 |
| 2012 | 22 | 6 | 48 | 24 |
| 2013 | 19 | 5 | 52 | 24 |
| 2014 | 20 | 4 | 46 | 30 |
| 2015 | 20 | 5 | 44 | 31 |
| 2016 | 23 | 4 | 44 | 29 |
| 2017 | 24 | 3 | 45 | 28 |
| 2018 | 20 | 3 | 40 | 37 |
| 2019 | 28 | 5 | 42 | 26 |
| 2020 | 31 | 4 | 34 | 31 |

The quarterly distribution of catch from 2010-2019 is similar to recent years with the northern summer fishery in Q3 accounting for the greatest proportion of the total catch. In 2020 the proportion in quarter 3 is still the highest at $34 \%$ but is similar to the quarter 1 and quarter 4 catches which both account for $31 \%$ of the total.

Catches per ICES statistical rectangle are shown in Figures 8.4.2.1 to 8.4.2.4. It should be noted that these figures are a combination of official catches and ICES estimates and may not indicate the true location of the catches or represent the location of the entire stock. These data are based on catches reported by all the major catching nations and represents almost the entire ICES estimated catch.

- $\quad$ First quarter 2020 ( $322419 \mathrm{t}-31$ \%)

The distribution of catches in the first quarter is shown in Figure 8.4.2.1. The proportion of the fishery taken in quarter 1 has increased in 2020 with the Scottish and Irish pelagic fleets targeting mackerel in Divisions 6.a, 7.b and 7.j. Substantial catches are also taken by the Dutch owned freezer trawler fleet. The largest catches were taken in Division 6.a, as in recent years. An increase in catch from 4.a and 7.b Q1 was seen again in 2020. The Spanish fisheries also take significant catches along the north coast of Spain during the first quarter.

- $\quad$ Second quarter 2020 (43 011 t-4 \%)

The distribution of catches in the second quarter is shown in Figure 8.4.2.2. The quarter 2 fishery is traditionally the smallest and this was also the case in 2020. The most significant catches where those in Division 8.c and at the start of the summer fishery in northern waters by Icelandic, Norwegian and Russian fleets in Division 2.a.

- $\quad$ Third quarter $2020(356006 \mathrm{t}-34 \%$ )

Figure 8.4.2.3 shows the distribution of the quarter 3 catches. Large catches were taken throughout Divisions 2.a (Russian, Norwegian and Faroese vessels), 4.a (Norwegian, Scottish vessels), 5.a (Icelandic vessels).

- Fourth quarter 2020 ( 318077 t - 31 \%)

The fourth quarter distribution of catches is shown in Figure 8.4.2.4. The proportion of the catch taken in the fourth quarter has increased from 26 \% in 2019 to $31 \%$ in 2020. The summer fishery in northern waters has largely finished with very small catches reported from Division 2.a. The largest catches are taken by Norway and Scotland around the Shetland Isles.

ICES cannot split the reported mackerel catches into different stock components because there is no clear distinction between components upon which a split could be determined. Mackerel with a preference for spawning in the northeast area, including the North Sea, cannot presently be identified morphometrically or genetically (Jansen and Gislason, 2013). Separation based on time and area of the catch is not a precise way of splitting mackerel with different spawning preferences, because of the mixing and migration dynamics including inter-annual (and possibly seasonal) variation of the spawning location, combined with the post-spawning immigration of mackerel from the south-west where spawning ends earlier than in the North Sea.

### 8.4.3 Catch-at-Age

This catch in numbers relates to a total ICES estimated catch of 1039513 t . These figures have been appended to the catch-at-age assessment table (see Table 8.7.1.2).

Age distributions of commercial catch were provided by Denmark, England, Germany, Faroes, Iceland, Ireland, the Netherlands, Norway, Portugal, Russia, Scotland, Northern Ireland and Spain. There remain gaps in the age sampling of catches, notably from France (length samples were provided), Sweden, Lithuania and Poland.

Catches for which there were no sampling data were converted into numbers-at-age using data from the most appropriate fleets. Accurate national fleet descriptions are required for the allocation of sample data to unsampled catches.

The catch numbers at age show a number of strong year classes in this fishery. Over $80 \%$ of the catch in numbers in 2020 consists of 3 to 10-year olds with all year classes between 2010 and 2014 contributing over $10 \%$ to the total catch by number. The 2016 year-class was strong in the fishery in 2020 and accounts for $11 \%$ of the catch numbers at age. The 2015 year-class does not look as strong as the other year and represents $5 \%$ of the total. In 2020 there is an increase in the proportion of fish in the plus group. Fish at $12+$ represent $7 \%$ of the total which is an increase from $3 \%$ in 2019.

There is a small presence of juvenile (age 0) fish within the 2020 catch. As in previous years catches from Divisions 8.c and 9.a have contained a proportion of juveniles.

### 8.5 Biological Data

### 8.5.1 Length Composition of Catch

The mean length-at-age in the catch for 2020 are given in Table 8.5.1.1.
For the most common ages which are well sampled there is little difference to recent years. The length of juveniles is traditionally rather variable. The range of lengths recorded in 2020 for 0 group mackerel ( $177 \mathrm{~mm}-266 \mathrm{~mm}$ ) is similar to 2019 ( $172 \mathrm{~mm}-267 \mathrm{~mm}$ ) and higher than those in 2018 ( $162 \mathrm{~mm}-254 \mathrm{~mm}$ ) and $2017(131 \mathrm{~mm}-212 \mathrm{~mm})$. The rapid growth of 0 -group fish combined with variations in sampling between northern and southern areas will contribute to the observed variability in the observed size of 0 -group fish. Growth is also affected by fish density as indicated by a recent study which demonstrated a link between growth of juveniles and adults ( $0-4$ years) and the abundance of juveniles and adults (Jansen and Burns, 2015). A similar result was obtained for mature 3- to 8-year-old mackerel where a study over 1988-2014 showed
declining growth rate since the mid-2000s to 2014, which was negatively related to both mackerel stock size and the stock size of Norwegian spring spawning herring (Ólafsdóttir et al., 2015).

### 8.5.2 Weights at Age in the Catch and Stock

The mean weight-at-age in the catch for 2020 are given in Table 8.7.1.3. There is a trend towards lighter weight-at-age for the most age classes (except 0 to 2 years old) starting around 2005, continuing until 2013 (Figure 8.5.2.1). This decrease in the catch mean weight-at-age seems to have stopped since 2013 and values for the last six years do not show any particular trend for the older ages (age 6 and older) and are slightly increasing for younger ages (ages 1 to 5). These variations in weight-at-age are consistent with the changes noted in length in Section 8.5.1.

The Working Group used weight-at-age in the stock calculated as the average of the weight-atage in the three spawning components, weighted by the relative size of each component (as estimated by the 2019 egg survey for the southern and western components and the 2017 egg survey for the North Sea component). Mean weight-at-age in 2020 for the western component are estimated from Dutch, Irish and German commercial catch data, the biological sampling data taken during the egg surveys and during the Norwegian tagging survey. Only samples corresponding to mature fish, coming from areas and periods corresponding to spawning, as defined at the 2014 benchmark assessment (ICES, 2014) and laid out in the Stock Annex, were used to compute the mean weight-at-age in the western spawning component. For the North Sea spawning component, mean weight-at-age in 2020 were calculated from samples of the commercial catches collected from Divisions $4 . a$ and $4 . b$ in the second quarter of 2020. Stock weights for the southern component, are based on samples from the Spanish catch taken in Divisions 8.c and 9.a in the $2^{\text {nd }}$ quarter of the year. The mean weights in the three component and in the stock in 2020 are shown in the text table below.

As for the catch weights, the decreasing trend observed since 2005 for fish of age 3 and older seems to have stopped in 2013 and values in the last 7 years do not show any specific trend (except for weights of ages 2 to 5 which have been increasing, Figure 8.5.2.2).
$\left.\begin{array}{lllll}\hline & \text { North Sea Component } & \text { Western } \\ \text { Age } & \text { Component }\end{array}\right)$

|  | North Sea Component | Western <br> Component | Southern Component |
| :--- | :--- | :--- | :--- | :--- | | NEA Mackerel |
| :--- |
| Age |

* Missing value of mean weight-at-age per component are replaced by component mean value in the calculation of the stock weights


### 8.5.3 Natural Mortality and Maturity Ogive

Natural mortality is assumed to be 0.15 for all age groups and constant over time.
The maturity ogive for 2020 was calculated as the average of the ogives of the three spawning components weighted by the relative size of each component calculated as described above for the stock weights. The ogives for the North Sea and Southern components are fixed over time. For the Western component the ogive is updated every year, using maturity data from commercial catch samples from Germany, Ireland, the Netherlands and the UK collected during the first and second quarters (ICES, 2014 and Stock Annex). The 2020 maturity ogives for the three components and for the mackerel stock are shown in the text table below.

| Age | North Sea <br> Component | Western <br> Component | Nouthern <br> Component | Mackerel |
| :--- | :--- | :--- | :--- | :--- |


| Age | North Sea | Western |
| :--- | :--- | :--- | :--- | :--- |
| Component | Component | Southern |
| Component |  |  |$\quad$| NEA |
| :--- |
| Mackerel |

A trend towards earlier maturation (increasing proportion mature at age 2) has been observed from around 2008 to 2015. A change in the opposite direction has been observed since then and the proportion of fish mature at age in 2020 are now markedly lower than in the previous years and at levels comparable with the ones observed at the end of the 2000s (Figure 8.5.3.1).

### 8.6 Fishery Independent Data

### 8.6.1 International Mackerel Egg Survey

### 8.6.1.1 Survey Planning for the 2022 Northeast Atlantic survey

The last mackerel egg survey (MEGS, I4189) was carried out in the NEA mackerel spawning areas in 2019 and a presentation with the final results were given during the WGWIDE meeting by the survey coordinator in 2020 (ICES, 2020a).

The ICES Working Group on Mackerel and Horse Mackerel Egg Surveys (WGMEGS) met in an online meeting in April 2021 to plan the international mackerel and horse mackerel egg survey in 2022. The nations participating in the 2022 MEGS survey will be Portugal, Spain, Scotland, Ireland, The Netherlands, Germany, Norway and the Faroe Islands.

In 2022, the MEGS survey in the western and southern areas for mackerel will continue as an Annual Egg Production Method (AEPM) survey; however, as with the surveys in 2013, 2016 and 2019, the intention will be to also carry out intensive Daily Egg Production Method (DEPM) adults sampling during the expected peak spawning period, in an attempt to calculate a DEPM SSB estimate.

WGMEGS considered a proposal to move the timing of the North Sea survey to the same year as the western surveys. If approved this survey would now be conducted by Denmark and England in 2022. Their participation would not lead to any reduction of available effort for the western surveys in 2022. This North Sea survey will be conducted as a DEPM survey (ICES, 2021).

The provisional survey plan of the 2022 mackerel and horse mackerel egg survey in the western and southern areas, as agreed during last the WGMEGS meeting (ICES, 2021), is presented in Table 8.6.1.1.1.

In preparation for the 2022 survey a workshop on Mackerel, Horse Mackerel and Hake Egg Identification and Staging (WKMACHIS) will take place during October 2021 and a Workshop on Adult Egg Production Methods Parameters estimation in mackerel and horse mackerel (WKAEPM) will be held in November 2021.

### 8.6.1.2 Changing from the Annual to Daily Egg Production Method.

From the start in 1977, WGMEGS has used the AEPM for estimation of NEA mackerel SSB (Lockwood et al. 1981; Lockwood, 1988) under the assumption that mackerel has a determinate fecundity. These surveys are carried out triennially.

The key concept for egg production method is very simple; if we know how many eggs have been spawned over a period of time (e.g., daily or annually) in the spawning area (egg production), and we know how many eggs an average individual mature female can produce over the
same period (fecundity), then we can estimate the size of the spawning population (Bernal et al., 2012).

There are two primary egg production methods (Gunderson, 1993; Hunter and Lo, 1993), namely the AEPM and the DEPM. The first method is designed for species with a determinate fecundity, i.e., those in which all the eggs to be spawned during the year are present and identifiable in the ovary immediately prior to spawning. With the AEPM, estimated total egg production is integrated over the whole annual spawning season and how many eggs are produced on average by female in the year (Costas et al., WD04 in Annex 05). Whereas the application of AEPM is suitable for determinate annual spawners, the DEPM can in principle be applied to indeterminate and determinate spawners.

The AEPM requires several ichthyoplankton surveys covering the whole spawning season and spawning area to estimate total annual egg production and sampling of pre-spawning adults to estimate annual potential fecundity. (Armstrong et al., 2012). Species with determinate fecundity have as an assumption that the fecundity is fixed before the onset of spawning (Hunter et al., 1992).

The DEPM can be used for species with an indeterminate fecundity, in which the potential fecundity is not fixed before the onset of spawning (Stratoudakis et al., 2006) and oocytes are recruited over the spawning season. The DEPM requires a single ichthyoplankton survey covering the entire spawning area during a brief period at or near the annual peak of spawning to estimate the mean daily egg production and to have representative samples of spawning adults during this survey period to estimate the mean daily fecundity (Parker, 1980; Stratoudakis et al., 2006). Accordingly, the DEPM provides a snapshot rather than an integrated view of the spawning season as the AEPM (Stratoudakis et al., 2006).

The main difference of the DEPM in relation to the AEPM method resides on the appropriate measure of fecundity, (Stratoudakis et al., 2006, Bernal et al., 2012).

In 2012, WGMEGS coordinated the Workshop on Survey Design and Mackerel and Horse Mackerel Spawning Strategy (ICES, 2012) as there are some indications that mackerel would be rather an indeterminate spawner and the DEPM might be more appropriate (Armstrong and Witthames, 2012). This workshop recommended that extra adult samples should be collected on surveys to investigate the estimation of DEPM adult parameters, and to attempt a contrast between AEPM and DEPM results.

During its 2018 WGMEGS meeting, after assessing the quality of the 2017 North Sea survey results, it was decided to consider utilizing DEPM for this survey, starting in 2020 (Costas et al., WD04 in Annex 05). Utilizing DEPM for the North Sea mackerel egg survey would have the advantage of requiring only one full coverage of the spawning area over a shorter time period (ICES, 2018b).

For the western and southern areas WGMEGS continues the use of the AEPM for mackerel.

### 8.6.1.3 2021 North Sea mackerel egg survey

The North Sea Mackerel Egg Survey (NSMEGS, I1582) is designed to estimate the spawning stock biomass (SSB) of mackerel of the North Sea spawning component of the Northeast-Atlantic stock on a triennial basis. Prior to 2017 this survey was done utilizing the AEPM. In the 2018 WGMEGS meeting, it was agreed to switch to the DEPM for the NSMEGS in 2020 (ICES, 2018b). However, due to the pandemic and the implementation of Covid-19 measures, the survey has to be postponed to 2021 (van Damme et al., WD01).The NSMEGS was carried out from 25th May to 12th June by The Netherlands, Denmark and Scotland. During this period the spawning area between $53^{\circ} \mathrm{N}$ and $62^{\circ} \mathrm{N}$ in the North Sea was covered by a total of 294 plankton stations and 22
pelagic trawl hauls were performed for the collection of mackerel adult and ichthyoplankton samples (Figure 8.6.1.3.1).

The spatial egg production distribution is shown in Figure 8.6.1.3.2. The mean Daily egg production was calculated for the total investigated area (Table 8.6.1.3.1).

The Netherlands sampled 524 mackerel during the survey and collected ovary samples of 164 females. Denmark sampled 817 mackerel during the survey and collected ovary samples of 119 females. The adult parameters are still very preliminary and without adult parameters the SSB cannot be estimated. When final fecundity parameter estimates are available and agreed by WGMEGS, an estimate of SSB will be provided to WGWIDE.

### 8.6.1.4 Results of the 2021 Exploratory Egg Survey in the Norwegian Sea.

Since 2007 WGMEGS has been observing and reporting on the offshore westwards and northwards expansion of NEA mackerel spawning. Initially spawning densities within these expanded areas were low, however the results from the most recent MEGS surveys in 2016 and 2019 provided clear evidence of a significant and unprecedented shift north and also westwards with some of the highest spawning densities observed being very close to the northern and north-western survey boundaries. During the last NEA mackerel benchmark in 2017 (ICES, 2017b) WGMEGS committed to undertake exploratory ichthyoplankton surveys within these remote boundary regions in the North and Northwest.

In 2017 and 2018 exploratory surveys undertaken by Ireland and Scotland as well as additional samples collected using existing Nordic surveys successfully mapped and delineated a mackerel spawning boundary within the North and northwest areas of Hatton Bank/South Iceland Basin and the Scotland-Faroe-Iceland Ridge (ICES, 2018b). The results and knowledge gleaned, informed the survey planning process ahead of the 2019 MEGS triennial survey but left the Norwegian Sea as an area that still provided a level of uncertainty and with the 2019 MEGS survey results providing evidence that mackerel appeared to be taking the North-eastern route towards their summer feeding grounds (Figure 8.6.1.4.1). A third and final exploratory survey was completed between the 7th - 22nd June 2021, (Burns and O' Hea, WD 15 in Annex 05) using the charter vessel Altaire. This would conclude the exploratory objective by surveying mackerel spawning activity up and along the Norwegian Sea and during the month when the highest mackerel spawning densities were likely to be encountered within this region. Additionally, 3 survey transects were also undertaken within the Northern North Sea area extending the survey's geographical footprint up to nearly 62 N .

78 plankton deployments were completed with the Gulf VII sampler during the survey, which due to the relatively calm conditions experienced throughout was able to survey as far North as Lofoten at 68.25 N . 5123 mackerel eggs of all stages were recorded during the survey, of which 1671 were recently spawned stage 1 eggs. Mackerel eggs were recorded from every deployment with stage 1 eggs being recorded on all but 2 of the stations completed. The numbers of mackerel eggs extracted from the Gulf VII samples were standardised and the stage 1 data presented as numbers $/ \mathrm{m}^{2} /$ day (Figure 8.6.1.4.2). Egg counts recorded during the survey area were generally low with the highest egg counts generally being reported within the southern half (south of 66N) of the survey area. Densities reduced gradually with increasing latitude until down to single figures on transects West of Lofoten as even surface temperatures approached the temperature threshold for spawning mackerel at between $8-9$ degrees Celsius. 2 successful deployments were completed with the vessels own midwater trawl providing 123 adult mackerel which were sampled for biological parameters and in addition 60 ovaries were also collected to progress ongoing research for IMR, Bergen.

Additional complementary plankton samples were collected by the Faeroe Islands during the IESNS survey during May 2021 and within the region extending from the east side of Iceland
across to the north of Faroe and Shetland. These samples were collected using a vertically deployed WP2 net that is lowered to a depth of 50 m . These samples have yet to be analysed but the results will be available prior to WGMEGS in 2022 and incorporated into the WG report.

The exploratory survey was unable to find a hard spawning boundary at its Northern extent albeit the numbers being encountered were very low at those high latitudes. This survey contrasted markedly with the previous exploratory surveys undertaken during 2017 and 2018 where the results reaffirmed the existence of the cold water barrier stretching from the East coast of Iceland across to the Faroe/Shetland channel and above which virtually no mackerel spawning takes place in June. The situation up and along the Norwegian Sea is very different with the influence of the Norwegian Current keeping sea surface temperatures (even at those high latitudes) well within a range that is tolerable for spawning mackerel. Nevertheless, the spawning levels observed in the sampled stations North of 62 degrees are overall very low with an estimated contribution to the overall total annual egg production (TAEP) of around $2-3 \%$. Looking ahead to the 2022 survey, WGMEGS therefore does not identify any immediate requirement to significantly extend the survey coverage in this region much beyond what was undertaken in 2019. All the information gathered from these exploratory egg surveys as well as the additional samples received from the various Nordic surveys since 2017 have proved to be invaluable and provide an opportunity not available during the triennial survey year to map the distribution of spawning mackerel within these remote northern boundary regions ahead of the triennial survey in 2022.

### 8.6.2 Demersal trawl surveys in October - March (IBTS Q4 and Q1)

## The data and the model

An index of survivors in the first autumn-winter (recruitment index) was derived from a geostatistical model fitted to catch data from bottom trawl surveys conducted during autumn and winter. A complete description of the data and model can be found in Jansen et al. (2015) and the NEA mackerel Stock Annex.

The data were compiled from several bottom trawl surveys conducted between October and March from 1998-2021 by research institutes in Denmark, England, France, Germany, Ireland, Netherlands, Norway, Scotland and Sweden. Surveys conducted on the European shelf in the first and fourth quarters are collectively known as the International Bottom Trawl Survey (IBTS), although several of the surveys use different names. All surveys sample the fish community on the continental shelf and upper shelf slope. IBTS Q4 covers the shelf from the Bay of Biscay to North of Scotland, excluding the North Sea, while IBTS Q1 covers the shelf waters from north of Ireland, around Scotland, the North Sea, Skagerrak and Kattegat.

Trawl operations during the IBTS have largely been standardized through the relevant ICES working group (ICES, 2013). Furthermore, the effects of variation in wing-spread and trawl speed were included in the model (Jansen et al., 2015). Trawling speed was generally $3.5-4.0$ knots, and trawl gear is also standardized and collectively known as the Grande Ouverture Verticale (GOV) trawl. Some countries use modified trawl gear to suit the particular conditions in the respective survey areas, although this was not expected to change catchability significantly. However, in other cases, the trawl design deviated more significantly from the standard GOV type, namely the Spanish BAKA trawl, the French GOV trawl, and the Irish mini-GOV trawl. The BAKA trawl had a vertical opening of only $2.1-2.2 \mathrm{~m}$ and was towed at only 3 knots. This was considered substantially less suitable for catching juvenile mackerel and, therefore, was excluded from the analysis. The French GOV trawl was rigged without a kite and typically had a reduced vertical opening, which may have reduced the catchability of pelagic species like mackerel. Catchability was assumed to equal the catchability of the standard GOV trawl because testing has shown that the recruitment index was not very sensitive to this assumption (Jansen et al.,
2015). Finally, the Irish mini-GOV trawl, used during 1998-2002, was a GOV trawl in reduced dimensions which was accounted for by inclusion of the wing-spread parameter in the model.

All surveys in 2020 Q4 and 2021 Q1 were conducted according to standards. Figure 8.6.2.1 provides an overview of the distribution and number of samples.

A geostatistical log-Gaussian Cox process model (LGC) with spatiotemporal correlations was used to estimate the catch rates of mackerel recruits through space and time.

## Results

The index of survivors in the first autumn-winter (recruitment index) was updated with data from surveys in 2020 Q4 and 2021 Q1. Parameter estimates and standard errors in the final model are listed in Table 8.6.2.1. An overview of the IBTS survey is given in Figure 8.6.2.1. The modelled average recruitment index (squared CPUE) surfaces were mapped in Figure 8.6.2.2a and b. The time series of spatially integrated recruitment index values is used in the assessment as a relative abundance index of mackerel at age 0 (recruits). All annual index values were estimated to be slightly higher than during the previous model fit (IBPNeaMAC: ICES, 2019), but with the same interannual pattern ( $p<0.001, r>0.99$ ). This increase does not affect the stock assessment because it is used in the assessment as a relative abundance index. The estimated index value for the 2020 year-class is above average (Figure 8.6.2.3).

## Discussion

The combined demersal surveys have incomplete spatial coverage in some areas that can be important for the estimation of age-0 mackerel abundance, namely: (i) Since 2011, the English survey (covering the Irish sea and the central-eastern part of the Celtic sea including the area around Cornwall) has been discontinued, (ii) the Scottish survey has not consistently covered the area around Donegal Bay, (iii) the IBTS has observed high catch rates in some years at the northeastern edge of the survey area (towards the Norwegian trench) in winter. It is therefore possible that some recruits are also overwintering on the other side of the trench along the south western shelf edge of Norway. Consequently, the NS-IBTS in Q1 should be extended to include the southwestern Norwegian shelf and shelf edge in proximity to the Norwegian trench.

Finally, WGWIDE encourages studies of vertical distribution and catchability of age-0 mackerel in the Q4 and Q1 surveys, to evaluate if it is comparable in all areas (see acoustic information in Jansen et al., 2015).

### 8.6.3 International Ecosystem Summer Survey in Nordic Seas (IESSNS, A7806)

IESSNS is the only annual survey providing data used in the assessment and covers summer feeding distribution of mackerel age 3+ in Nordic Seas. In 2021, survey coverage in the western area was reduced as Greenlandic waters, Iceland basin (south of latitude $62^{\circ} 45^{\prime}$ ) and the Reykjanes ridge (south of latitude $62^{\circ} 45^{\prime}$ ) were not surveyed. Coverage reduction did no impact quality of the survey as zero mackerel boundary was established north, west, and south of Iceland. The survey was successfully conducted in 2021. IESSNS cruise report is available as a working document to this report and a detailed survey description is available in the mackerel Stock Annex.

Abundance estimates by age are displayed in input data for the assessment (Table 8.7.1.9), survey estimates of total stock abundance and stock biomass with confidence intervals in Figures 8.6.3.12, internal consistency of mackerel abundance from 2012 to 2021 is displayed in Figure 8.6.3.3 and catch curves abundance at age from 2010 to 2021 in Figure 8.6.3.4. Estimated total stock abundance and total biomass declined $53 \%$ and $58 \%$ respectively compared to 2020 . Abundance
declined for all cohorts age 3+ but the decline was greater for age $5+$. Internal consistency declined compared to 2020, particularly for ages $5-8$ years. This is a sudden and unexpected decline in mackerel abundance compared to 2019-2020 but when compared to 2018 it is $28 \%$ lower. Further analysis of the IESSNS time series is needed to evaluate if the survey index is an overestimate in 2019-2020 or an underestimate in 2018 and 2021. The sudden drop in abundance is reflected in declining internal consistency and drop in catch curves. Bootstrap estimation of abundance by age displayed in Figure 8.6.3.5. Swept area trawl catch and mean catch rate for 2021 is displayed in Figure 8.6.3.6 and mean mackerel catch rate per rectangle for years 2010 and from 2012 to 2021 in Figure 8.6.3.7.

### 8.6.4 Tag Recapture data

The following is a summary of the most important information on tag recapture data, more detailed info can be found in a working document attached to this report (Slotte and Hølleland, WD06 in Annex 05). Information from steel tagging experiments conducted by Institute of Marine Research in Bergen (IMR) on mackerel at spawning grounds west of Ireland and British Isles in May-June and the respective recaptures at Norwegian factories with metal detectors (Tenningen et al., 2011) was introduced to the mackerel assessment during ICES WKPELA 2014 (ICES, 2014). Data from release years 1980-2004, and recapture years 1986-2006 have been used in the update assessments following this benchmark. From 2011 onwards IMR changed tagging methodology to radio-frequency identification (RFID), more specifically passive integrated transponder tags (PIT-tags). This allowed for more automatic data processes with recaptures from scanned landings at factories in Norway, Scotland and Iceland now being updated real time in an IMR data base over internet.

The data format is the same for both tag types; a table showing numbers of tagged fish per year class in each release year, and the corresponding numbers scanned and recaptured of the same year classes in all years after release. The RFID data were considered to be a new time series with a different scaling factor (survival) than the steel tags, and it has been used in update assessments following the ICES WKWIDE2017 benchmark (ICES, 2017). For steel tags data from ages 2-11 and all recapture years are used in the assessment. During the 2017 benchmark it was decided to use the same filtering for the RFID data from release year 2011 onwards. However, following decisions made during ICES IBPNEAMac 2019 (ICES, 2019) update assessments are now only using RFID data from release years 2013 onwards, ages 5-11 and recapture year 1 and 2 after release.

An overview of all RFID tagging data in terms of numbers tagged, biomass scanned, and numbers recaptured per year, and geographical distributions of data are shown in Figures 8.6.4.1-3. The exclusion of recapture years 3 and longer after release is due to potential tag loss over time, which seem evident in the RFID data (Slotte and Hølleland, WD06 in Annex 05). The exclusion of release years 2011-2012 is mainly based in lack of distributional coverage of scanned fishery, which changed significantly when more countries joined the program from 2014 onwards (Figure 8.6.4.2). The exclusion of ages 1-4, was mainly based on the fact that early in the time series these age groups were relatively few compared with the scanned fish year 1 and 2 after release, leading to some noise in the data. However, the age structure of tagged and scanned fish year 12 after release has developed over time series to be more overlapping, and high proportions of tagged mackerel are now at ages 2-4 (Figure 8.6.4.4).

Trends in year class abundance indices from RFID data based on recaptures year 1 and 2 after release now seem consistent and informative for assessment from ages 2-12 (Figure 8.6.4.5). Note that an alternative assessment at WGWIDE2021 using these indices for the selected ages 5-11 instead of the regular data table resulted in negligible differences in SSB trend and same leave out RFID data effects; i.e., higher SSB in most recent years when excluding RFID data. Translating
these abundance indices into different age-aggregated biomass indices also show comparable time trend with SSB from WGWIDE2021 from release years 2013 onwards (Figure 8.6.4.5). Especially the marked decrease in SSB from 2017-2019 seem to follow the decline in the RFID biomass estimates, which may explain why leave out RFID runs from WGWIDE2021 tends to lift the SSB upwards. The signals of total mortality rate ( Z ) in fully mature fish aged 4-12 for year classes 2003-2014 tends to be higher in the RFID data than in the catch data tightly overlapping with Z signals in the final WGWIDE2021 assessment, whereas for the international trawl survey IESSNS the estimated Z is even lower (Figure 8.6.4.6).

The overall conclusion is that the RFID time series is slowly developing, but still is a very short time series. Nevertheless, the data seem quite informative for stock assessment, although showing higher total mortality rate signals than the other input data. Such conflicting trends suggest that year to year variations in assessment and leave out effects may frequently occur in coming years when time series are short. Finally, the new development of the time series suggests that the current filtering of RFID data for use in stock assessment should be revised in near future. This especially counts for the inclusion of younger ages 2-4 that may be informative for incoming year classes to the stock.

### 8.6.5 Other surveys

### 8.6.5.1 International Ecosystem survey in the Norwegian Sea (IESNS, A3675)

After the mid-2000s an increasing amount of NEA mackerel has been observed in catches in the Norwegian Sea during the combined survey in May during the International Ecosystem survey in the Norwegian Sea (IESNS) targeting herring and blue whiting (Salthaug et al. 2019; 2020). The spatial distribution pattern of mackerel was quite similar in 2020 compared to 2019 Salthaug et al., 2019). Mackerel was caught within a more expended area and in more trawl stations of the Norwegian Sea in May 2020 compared to May 2019 (Salthaug et al., 2019; 2020). In 2020, the northernmost mackerel catch was at $69^{\circ} \mathrm{N}$ and the westernmost catch was around $4^{\circ} \mathrm{W}$, which is further north and west than recorded in 2019 (Salthaug et al. 2019; 2020). Mackerel of age 4 dominated, followed by age 6 in 2020, whereas there was found more 1-year olds compared to last year, particularly in the north (Salthaug et al., 2020). Mackerel was present in the southern and eastern part of the Norwegian Sea (as far north as $68^{\circ} \mathrm{N}$ ) in the beginning of May 202I.

The IESNS survey provides valuable, although limited, quantitative information on mackerel. This acoustic based survey is not designed to monitor mackerel, and does not provide proper mackerel sampling in the vertical dimension and involves too low trawl speed for representative sampling of all size groups of mackerel. The trawl hauls are mainly targeting acoustic registrations of herring and blue whiting during the survey in May (IESNS) (Salthaug et al., 2019, 2020, WD14 in Annex 05). Therefore, no further quantitative information can be drawn from these data as this survey is not designed to monitor mackerel.

### 8.6.5.2 Acoustic estimates of mackerel in the Iberian Peninsula and Bay of Biscay (PELACUS, A2548)

PELACUS survey data have not been processed on time for WGWIDE and therefore, no new information from the Bay of Biscay on mackerel distribution and abundance during spawning time is available.

### 8.7 Stock Assessment

### 8.7.1 Update assessment in 2021

The update assessment was carried out by fitting the state-space assessment model SAM (Nielsen and Berg, 2014) using the R library stockassessment (downloadable at install_github("fishfollower/SAM/stockassessment")) and adopting the configuration described in the Stock Annex.

The assessment model is fitted to catch-at-age data for ages 0 to 12 (plus group) for the period 1980 to 2020 (with a strong down-weighting of the catches for the period 1980-1999) and three surveys: 1) the SSB estimates from the triennial Mackerel Egg survey (every three years in the period 1992-2019); 2) the recruitment index from the western Europe bottom trawl IBTS Q1 and Q4 surveys (1998-2020); and 3) the abundance estimates for ages 3 to 11 from the IESSNS survey (2010, 2012-2021). The model also incorporates tagging-recapture data from the Norwegian tagging program (for fish recaptured between 1980 and 2005 for the steel tags time series, and fish recaptured between 2014 and 2020 (age 5 and older at release) for the radio frequency tags time series).

Fishing mortality-at-age and recruitment are modelled as random walks, and there is a process error term on abundances at ages 1-11.

The differences in the new data used in this assessment compared to the last year's assessment were:

- Update of the recruitment index until 2020.
- Addition of the 2021 survey data in the IESSNS indices.
- Addition of the 2020 catch-at-age, weights-at-age in the catch and in the stock and maturity ogive, proportions of natural and fishing mortality occurring before spawning.
- Update of the catch-at-age and mean-weight-at-age in the catch for the period 2010-2019 (see Section 8.4.3).
- The inclusion of the tag recaptures from 2020.

Input parameters and configurations are summarized in Table 8.7.1.1. The input data are given in Tables 8.7.1.2 to 8.7.1.9. Given the size of the tagging data base, only the data from the last year of recaptures is given in this report (Table 8.7.1.10).

### 8.7.2 Model diagnostics

## Parameter estimates

The estimated parameters and their uncertainty estimates are shown in Table 8.7.2.1 and Figure 8.7.2.1. The model estimates different observation standard deviations for young fish and for older fish. Reflecting the suspected high uncertainty in the catches of age 0 fish (mainly discards), the model gives a very poor fit to this data (large observation standard deviation). The standard deviation of the observation errors on catches of age 1 is lower, though still high, indicating a better fit. For the age 2 and older, the fit to the catch data is very good, with a very low observation standard deviation.

The observation standard deviations for the egg survey and the IESSNS surveys ages 4 to 11 are higher indicating that the assessment gives a lower weight to the information coming from these surveys compared to the catches. The IESSNS age 3 is very poorly fitted in the assessment (high observation standard deviation). Overdispersion of the tag recaptures has the same meaning as the observation standard deviations, but is not directly comparable.

The catchability of the egg survey is 1.22 , larger than 1 , which implies that the assessment considers the egg survey index to be an overestimate. The catchabilities at age for the IESSNS increase from 0.81 for age 3 to 1.95 for age 7 and 9 . Since the IESSNS index is expressed as fish abundance, this also means that the assessment considers the IESSNS to provide over-estimated abundance values for the oldest ages. The post tagging mortality estimate is higher for the steel tags (around $40 \%$ ) than for the RFID tags (around $15 \%$ ).

The process error standard deviation (ages 1-11) is moderate as well as the standard deviation of the F and recruitment random walks.

The catchability parameters for the egg survey, recruitment index and post tagging survival appear to be estimated more precisely than other parameters (Table 8.7.2.1). The catchability for the IESSNS have a slightly higher standard deviation, except for the catchability of the IESSNS at age 3 which has a much higher standard deviation. Uncertainty on the observation standard deviations is larger for the egg survey, the IESSNS age 3 , for the recruitment index and for the catches at age 1 than for the other observations. The uncertainty on the observation variance estimates is not particularly high, especially for the data sources with the lowest observation variances, which are the most influential on the assessment (Figure 8.7.2.2). Uncertainty on the overdispersion of the RFID tag data is high. The standard deviation on the estimate of process error is low, and the standard deviations for the estimates of F random walk variances of age 0 and 1 are both very high. The uncertainty on the random walk variance for recruitment is very large, indicating that the parameter was poorly estimated.

The estimated AR1 error correlation structure for the observations from the IESSNS survey age 3 to 11 has a high correlation between the errors of adjacent ages ( $r=0.77$ ), then decreasing exponentially with age difference (Figure 8.7.2.3.). This high error correlation implies that the weight of this survey in the assessment is lower than for a model without correlation structure, which is also reflects in the high observation standard deviation for this survey.

There are some correlations between parameter estimates (Figure 8.7.2.4):

- Catchabilities are positively correlated (especially for the IESSNS age 4 to 11), and negatively correlated to the survival rate for the RFID tags. This simply represents the fact that all scaling parameters are linked, which is to be expected.
- The observation variance for the recruitment index is inversely correlated to the variance of the random walk of the recruitment. This implies that when the model relies less on the recruitment index, the estimated recruitment time series becomes smoother.


## Residuals

The "one step ahead" (uncorrelated) residuals for the catches did not show any temporal pattern (Figure 8.7.2.5) except for 2014 for which they were mainly positive for 2014 (modelled catches lower than the observed ones). This may result from the random walk that constraints the variations of the fishing mortality, which prevents the model from increasing the fishing mortality suddenly (which probably happened given the sharp increase in the catches in 2014). Residuals are of a similar size for all ages, indicating that the model configuration with respect to the decoupling of the observation variances for the catches is appropriate. Residuals for the 2020 catches-at-age show that the model was not able to fully reproduce the strong increase in the catches of fish of age 9 and older although the estimated fishing mortality on the older fish has increased substantially 2020 (see results in section below).

The residuals for the egg survey show a strong temporal pattern with large positive residuals for the period 2007-2010-2013, followed by large negative residuals in 2016 and 2019. This pattern reflects the fact that the model, based on all the information available, does not follow the recent trend present in the egg survey (with an historical low estimate for 2019) and considers those
two last years as large negative observation errors. The relatively high observation variance for this survey indicates a poor fit with the egg survey due mainly to these two observations which point towards a very different direction from the other observations. Residuals for the IESSNS indices are relatively well balanced for most of the years, except for the 2019 and 2020 index, where residuals tend to be mainly positive. Despite the strong drop in the abundances at age in 2021, the residuals for this year do not indicate any year effect (e.g., no large residuals of the same sign observed across ages) . Residuals to the recruitment index show no particular pattern, and appear to be relatively randomly distributed in the earlier years, but positive residuals are consistently observed over the last 5 years, indicating that the model has difficulties agreeing with this sustain period of high values in the index.

Finally, inspection of the residuals for the tag recaptures (Figure 8.7.2.6) did not show any specific pattern for the RFID data. For the steel tags, there is a tendency to have more positive residuals at the end of the period which could indicate that using a constant survival rate for this dataset may not be appropriate.

## Leave one out runs

In order to visualise the respective impact of the different surveys on the estimated stock trajectories, the assessment was run leaving out successively each of the data sources (Figure 8.7.2.7).

All leave one out runs showed parallel trajectories in SSB and Fbar, except the one leaving out the RFID tag information, which shows a less steep decline in SSB since 2014, and continued decline in Fbar in the most recent years. For recruitment, all runs also resulted in similar trajectories, expect the run without the recruitment index, which recruitment decreased from high levels in the mid-2010s to historical low levels currently.

Removing the IESSNS resulted in lower SSB estimates and higher Fbar estimates for the period covered by the survey. Removing the recruitment index had a similar effect on SSB and Fbar. On the opposite, removing the egg survey results in a larger estimated stock, exploited with a lower fishing mortality. In both cases, the estimated stock trajectories are well within the confidence interval of the assessment using all data sources. As in previous years, the update assessment seems to make a trade-off between the information coming from the IESSNS which leads to a more optimistic perception of the stock, and the information from the egg survey which suggests a more pessimistic perception of the stock. The run leaving out the RFID also resulted in a higher SSB than in the assessment using all data, and a slightly higher fishing mortality between 2007 and 2014, but higher after 2016. The magnitude of the effect of removing the RFID data is similar to removing other surveys. This is a contrasting situation compared to the 2020 WGWIDE assessment, in which the RFID had a very small influence on the assessment (no effect on estimated stock trajectory, slightly reduced uncertainty when RFID data are included). This indicates that the influence of the RFID data compared to other data sources has increased this year. This point is further discussed below in a section presenting additional exploratory runs (Section 8.7.5.2.).

## Additional sensitivity runs

A series of additional sensitivity runs were done to identify the cause of the change in stock trajectories in the 2021 WGWIDE assessment compared to previous years assessment (see Section 8.10 for a description of this revision).

First, the influence of revisions in the historical data (catch-at-age and mean weight-at-age in the catch for the years 2010-2019) was tested by running the assessment using last year's data for 2010-2019, but keeping the new 2020. This run was almost identical to the WGWIDE 2021 update assessment (not presented here).

Then, the influence of the data added in 2021 was tested by running the model removing separately each of the new data added in 2021 (2020 catch-at-age, 2020 recruitment index, 2020 RFID
recaptures and 2021 IESSNS index). The two model runs excluding the 2020 recruitment index and the 2021 IESSNS are very similar to the current assessment and are not shown on Figure 8.7.2.8.

The exclusion of the 2020 RFID data leads to larger SSB and lower Fbar estimates over the most recent years (2019-2020). The information from the 2020 recaptures indicate that abundance has declined in 2019 for the third year in a raw. Adding this information to the assessment therefore leads to the reduction of stock abundances, and hence SSB.

The 2020 catch-at-age also seem to have a strong influence on the assessment. Excluding this information leads to stock trajectories very similar to those from the WGWIDE2020 assessment. The stock trajectories are revised over almost a decade (since about 2009), with lower SSB and higher Fbar estimated when the 2020 catches are not used. The data for 2020 are characterised by a sharp increase in the catches for the older fish (age 9 and older, including the plus group) compared to 2019. No particular changes in fishing patterns for the fleets have been reported and the reason for this increase is not fully understood. Given the low observation variance for the catch-at-age 2 and older, the SAM model follows tightly this increase in the catches of 9+ fish in 2020. The fit to these higher catches can be achieved partly by increasing the fishing mortality on the older age. However, the extend by which fishing mortality-at-age can increase in a year is limited by the amplitude of the random walk, and the variance of these processes is rather low for the mackerel assessment (Table 8.7.2.1). In addition, to be able to fit these higher catches, the model estimated relatively large abundances for old fish in 2020, which seems to have caused an upward revision of the abundance of these cohorts as far back in time as 2014 (based on the comparison of abundance-at-age from last year's and this year's assessment, no shown). This upward revision for abundance-at-age explains the downwards revision of fishing mortality at age. Last year's assessment (WGWIDE 2020; ICES, 2020a) was also quite sensitive to addition of a latest year of catch data (analysis done this year and hence not presented in the previous report) but the sensitivity is larger this year, probably due to the unexpected catches of old fish.

### 8.7.3 State of the Stock

The stock summary is presented in Figure 8.7.3.1 and Table 8.7.3.1. The stock numbers-at-age and fishing mortality-at-age are presented in Tables 8.7.3.2-3. The spawning stock biomass is estimated to have increased almost continuously from just above 2 million tonnes in the late 1990s and early 2000s to 5.55 million tonnes in 2014 and 2015 and subsequently declined to reach a level just above 3.87 million tonnes in 2019 and increase slightly in 2020 to 3.94 million tonnes. The fishing mortality has declined from levels between $\mathrm{F}_{\mathrm{pa}}(0.36)$ and $\mathrm{F}_{\mathrm{lim}}(0.46)$ in the mid-2000s to levels well below $\mathrm{F}_{\text {MSY }}(0.26)$ since 2015 and increased to just under Fmsy in 2020. The recruitment time series from the assessment is not considered a reliable indicator of year-class strength (see Section 8.7.5.1).

There is some indication of changes in the selectivity of the fishery over the last 30 years (Figure 8.7.3.2.). In the years 1990 s, the fishery seems to have had a steeper selection pattern (more rapid increase in fishing mortality with age). Between the end of the 1990s and the end of the 2000s, the selection pattern became less steep (decreasing selection on the ages2-5). After 2008, the pattern changed again towards a steeper selection pattern.

### 8.7.4 Quality of the assessment

## Parametric uncertainty

Large confidence intervals are associated with the SSB in the years before 1992 (Figure 8.7.4.1 and Figure 8.7.2.7). This results from the absence of information from the egg survey index, the
down-weighting of the information from the catches and the assessment being only driven by the tagging data and natural mortality in the early period. The confidence intervals become narrower from the early 1990s to the mid-2000s, corresponding to the period where information is available from the egg survey index, the tagging data and (partially) catches. The uncertainty increases slightly in the most recent years and the SSB estimate for 2020 is estimated with a precision of $+/-24 \%$ (Figure 8.7.3.1 and Table 8.7.3.1). There is generally also a corresponding large uncertainty on the fishing mortality, especially before 1995. The estimate of Fbart-8 in 2020 has a precision of $+/-25 \%$.

## Model instability

The retrospective analysis was carried out for 7 retro years, (or peels) by fitting the assessment using the 2021 data, removing successively 1 year of data (Figure 8.7.4.2.). There was a systematic retrospective pattern found in Fbar for the older retrospective peels (current year -3 to current year -7) with a systematic downwards revision. However, this pattern is not apparent in the most recent peels, and the Mohn's rho value of the last 5 years is of 0.16 . There is no retrospective pattern in the SSB and the value of the Mohn's rho on SSB for the last 5 peels if low ( -0.03 ). Recruitment appears to be quite consistently estimated for the 6 older retrospective peels, but over the last 2 peels, recruitment has been revised downwards. This is related to the increase in the observation variance for the recruitment index, and corresponding decrease in recruitment random walk variance. Recruitment estimates have progressively become less influenced by the recruitment index (which displays high value in the recent years and revised recent estimates upwards).

## Model behaviour

The realisation of the process error in the model was also inspected. The process error expressed as annual deviations in abundances-at-age (Figure 8.7.4.3) shows indications of some pattern across time and ages. There is a predominance of positive deviations in the recent years for ageclasses 5 to 8 . While process error is assumed to be independent and identically distributed, there is clear evidence of correlations in the realisation of the process error in the mackerel assessment, which appears to be correlated both across age-classes and temporarily.

The temporal autocorrelation can also be visualised if the process error is expressed in term of biomass (process error expressed as deviations in abundances-at-age multiplied by weight at age and summed over all age classes, Figure 8.7.4.4). Periods with positive values (when the model globally estimates larger abundances-at-age than corresponding to the survival equation) have been alternating with periods with negative values (1991-1994 and 2004 and 2006). For the years between 2008 and 2017, the biomass cumulated process error remains positive, and large (reaching in 2013 almost the weight of the catches). The reason for this misbehaviour of the model could not be identified.

### 8.7.5 Exploratory runs

### 8.7.5.1 Assessment starting at age 2

The age 0 estimates in the current assessment mainly rely on the recruitment index; the catch-atage 0 information is considered by the mode as uninformative (large observation variance). Catch-at-age information becomes influential at age 2 (very low observation variance). The recruitment signal provided by abundances estimated at age 2 or 3 (when the fish enters the fishery), is different from the signal in the age 0 abundance (Figure 8.7.5.1). Age 0 abundances are less variable than abundances at age 2 and 3 . For the period before 2012, there is a broad agreement in the perception of year class strength, although some year classes that do not appear particularly large at age 0 are perceived as very large at age 2 and 3 (e.g., 2002 year-class). For the more recent period, there is a greater discrepancy between recruitment at age 0 and at older
ages. While the age 0 abundances indicate very high recruitment for the year-classes 2012 to 2018, number of those year-classes appear as particularly poor based on age 2 and 3 abundances (2015, 2017 and 2018). As very little fishing occurs between age 0 and 2 and 3, exploitation is not likely to explain these changes in perception of cohort strength. Such variations could be possibly due to variations in natural mortality (e.g., the strength of a cohort may not be fully determined at age 0 and processes occurring during the first years of life may still be determining year-class strength). However, processes occurring at the juvenile stage are more likely to dampen the variations in cohorts' size (e.g., density dependent mechanisms) than increasing it. In addition, some cohorts increase in size as they become older (e.g., 2001 and 2002), which clearly indicates that this is more likely a model artefact. The cohort strength at age 0 , based on the recruitment index, is progressively revised, thanks to the process error occurring on annual survival, so that cohort strength at age 2 corresponds to the information coming from the catches.

This discrepancy between the recruitment estimates at age 0 and the actual size of the cohort when entering the fishery implies that the age 0 recruitment does not give an accurate indication on year-class strength, and should not be used to make assumption on stock development in the near future. This has implications for the short term forecast done to compute the catch advice, in which last estimated recruitment value (R2020 this year) contributes to around $10 \%$ of the catch and SSB in the advice year.

As very little fishing occurs on 0 and 1 year olds, and catch-at-age data is considered very noisy, and since there appears to be a disagreement between the recruitment index at age 0 and at older ages in the recent years, it does not seem relevant to start the assessment at age 0 or 1 . An exploratory run was conducted starting the assessment at age 2 (and hence removing catch-at-age information for age 0 and 1 and the recruitment index, while leave the rest of the data and model configuration unchanged).

The estimated parameters had in general similar values in the 2 models (Table 8.7.5.1) with a largest difference of $6 \%$ for the IESSNS catchability at age 3, except for the process variances where large differences are observed. Recruitment variability increases by $246 \%$, and this is associated to an $80 \%$ decrease on the standard deviation (uncertainty) on this parameter. F random walk variance increase by $24 \%$ (with a $24 \%$ reduction on the standard deviation) and the process error variance is reduced by $16 \%$ (but this a larger standard deviation). The model starting at age 2 therefore gives a similar weight to the different data sources as the current model (same observation variances) but estimates a much more variable recruitment, and slightly more variable fishing mortality.
Both assessments give a very similar perception of the SSB and Fbar trajectories (Figure 8.7.5.2). There is a small different in SSB in the years 2010 and 2011, and in the last year with catch information (2020). Fbar trajectories are very consistent, with slightly larger variations for the assessment starting at age 2. The recruitment at age 2 (in blue on Figure 8.7.5.2, note that the curve should be shifted backwards by 2 years to compare year-class strength with the recruitment at age 0 , red curve) shows a much variable year-class strength signal, with the same perception of year class strength as the age 0 recruitment for some years (broadly between year-classes 2000 and 2012), but a much lower estimated year-class strength since 2012.

In conclusion, both models broadly agree both in terms of fit to the data and in terms of stock trajectories, and the model starting at age 2 could be considered as potential alternative to the current model at the next benchmark for this stock. The two models however have very different implications regarding advice. While the current model assumes a high 2020 year-class, that will contribute to $10 \%$ in the SSB and catch and advice year (age 2), the alternative model suggests a low 2018 year-class (age 4 in advice year) and average recruitments (geometric mean assumption) for the 2019 and 2020 year classes (age 3 and 2 in advice year).

### 8.7.5.2 Assessment using tag data as abundance indices

The last inter-benchmark (ICES, 2019) showed that the RFID tagging data had a very high influence on the previous assessment, simply due to the fact that it was a much larger dataset than other survey data (and growing much faster as well). The changes made during this IBP involved filtering out a large part of the RFID dataset (tags recovered after more than 2 years at liberty were excluded due to the suspicion of tag loss). At the time of the IBP, this decreased considerably the weight of the RFID data on the assessment (as measured then by the leave one out run). This year, with 2 additional years of data, the RFID dataset has grown by 28 data points, while the second largest index, the IESSNS, has grown by 18 data point. At the same time, the leave one out run (Figure 8.7.2.7) shows that the influence of the RFID dataset has increased markedly compared to last year. It is unclear whether this increasing influence is due to the RFID data being very informative, and therefore receiving a higher weight, or if it is due to the increase in the number of observations.

In order to investigate this, the SAM model was fitted using the RFID tag data expressed as abundance-at-age indices for the ages 5 to 11 (see Figure 8.6.4.5). In this configuration, the RFID data has a similar number of observations as for the IESSNS survey. The assessment using RFID as indices gives a perception of the stock very similar to the WGWIDE 2021 assessment (Figure 8.7.5.3). There is hardly any difference in the estimated SSB, and Fbar and recruitment are slightly higher. This strong similarity between the assessments using the RFID data as recaptures or as abundance indices indicates that the stronger influence of the RFID seen for the WGWIDE2021 is not likely to be due to the larger increase in number of data points compared to other data sources, but rather to the information contained in the dataset.

### 8.8 Short term forecast

The short-term forecast provides estimates of SSB and catch in 2022 and 2023, given assumption of the current year's (also called intermediate year) catch and a range of management options for the catch in 2022.

All procedures used this year follow those used in the benchmark of 2014 as described in the Stock Annex.

### 8.8.1 Intermediate year catch estimation

Estimation of catch in the intermediate year (2021) is based on declared quotas and interannual transfers as shown in the text table in Section 8.1.

### 8.8.2 Initial abundances at age

The recruitment estimate at age 0 from the assessment in the terminal assessment year (2020) was considered too uncertain to be used directly, because this year class has not yet fully recruited into the fishery. The last recruitment estimate is therefore replaced by predictions from the RCT3 software (Shepherd, 1997). The RCT3 software evaluates the historical performance of the IBTS recruitment index, by performing a linear regression between the index and the SAM estimates over the period 1998 to the year before the terminal year. The recruitment is then calculated as a weighted mean of the prediction from this linear regression based on the IBTS index value, and a time tapered geometric mean of the SAM estimates from 1990 to the year before the terminal year. The time tapered geometric mean gives the latest years more weight than a geometric mean. This is done because the recent productivity of the stock appears different than in the 1990's.

The weighting calculated by RCT3 was 76 \% (recruitment index) and 24 \% (time tapered geometric mean), which leads to an expected recruitment of 5743 million.

### 8.8.3 Short term forecast

A deterministic short-term forecast was calculated using FLR (www.flr-project.org). Table 8.8.3.1 lists the input data and Tables 8.8.3.2 and 8.8.3.3 provide projections for various fishing mortality multipliers and catch constraints in 2022.

Assuming catches for 2021 of 1199 kt , F was estimated at 0.35 (above Fmsy) and SSB at 3.51 Mt (above $\mathrm{B}_{\mathrm{pa}}$ ) in spring 2021. If catches in 2022 equal the catch in 2021, F is expected to increase to 0.42 (above $\mathrm{F}_{\mathrm{pa}}$ ) in 2022 with a corresponding decrease in SSB to 3.21 Mt in spring 2022. Assuming an F of 0.42 again in 2023, the SSB will further decrease to 2.89 Mt in spring 2023.

Following the MSY approach, exploitation in 2022 shall be at $\mathrm{F}_{\mathrm{mSy}}$ (0.26). This is equivalent to catches of 795 kt and a decrease in SSB to 3.31 Mt in spring 2022 ( $6 \%$ decrease). During the subsequent year, SSB will remain at a similar level (3.27 Mt) in spring 2022.

### 8.9 Biological Reference Points

A management strategy evaluation Workshop on northeast Atlantic mackerel (MKMSEMAC) was conducted during 2020 (ICES, 2020b) which resulted in the adoption of new reference points for NEA mackerel stock by ICES.

The table below summarises the currently used reference points.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 2.58 million tonnes | $\mathrm{B}_{\mathrm{pa}}$ | ICES (2020b) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.26 | Stochastic simulations | ICES (2020b) |
| Precautionary approach | $\mathrm{Blim}^{\text {lim }}$ | 2.00 million tonnes | Bloss in 2003 from the 2019 WGWIDE assessment (ICES, 2019) | ICES (2020b) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 2.58 million tonnes | $B_{\lim } \mathrm{x} \exp (1.645 \times \sigma)$, with $\sigma_{S S B}=0.15$ | ICES (2020b) |
|  | $F_{\text {lim }}$ | 0.46 | F that, on average, leads to $\mathrm{Blim}_{\text {lim }}$ | ICES (2020b) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.36 | $\mathrm{F}_{\mathrm{p} 05}$ (the F that leads to SSB $\geq$ Blim with 95 \% probability) | ICES (2020b) |

### 8.10 Comparison with previous assessment and forecast

## Stock assessment output

The last available assessment used for providing advice was carried out in 2020 during the WGWIDE. The new 2021 WGWIDE assessment gives a slightly different perception of the development of the stock, with a higher SSB estimated for the period 2014-2017 and a lower Fbar estimated over the period 20092018 (Figure 8.10.1). For the latest year, the differences in the 2019 TSB, SSB and Fbar estimates between the previous and the present assessments are small, of $0.7 \%, 3.9 \%$ and $-3.6 \%$, respectively. The 2018 fishing mortality is unchanged ( $0.2 \%$ difference).

|  | TSB 2019 | SSB 2019 | Fbar4-8 2019 |
| :--- | :--- | :--- | :--- |
| Values |  |  |  |
| 2020 WGWIDE | 4966328 tonnes | 3731510 tonnes | 0.223 |
| 2021 WGWIDE | 4933409 tonnes | 3876306 tonnes | 0.215 |
| \% difference | $-0.7 \%$ | $3.9 \%$ | $-3.6 \%$ |

The addition of a new year of data has slightly modified model parameters compared to last year (Figure 8.10.2). The observation standard deviation has decreased for the IESSNS survey, and increased for the egg survey (although changes are very minimal in both cases). The observation standard deviation for the recruitment index increased by a larger proportion. This increase comes with a substantial decrease of the random walk variance for recruitment, and a larger uncertainty on this parameter. The 2021 model fit follows less the recruitment index and, in absence of other source of information on age 0 , produces a smoother recruitment time series.

Although the parameters corresponding to the weight of the different data sources on the assessment (observation standard deviations) have not changed, the analyses presented in Section 8.7 indicated that the influence of the RFID time series has increased. In addition, Section 8.7 also showed that the revision observed this year is mainly due to the influence of the inclusion of the 2020 catch at age, which effect propagated backward in time.

The uncertainty on the parameter estimates has decreased for some parameters (observation standard deviation on the IESSNS survey, standard deviations of the F random walk for age 0 and 1, figure 8.10.2), but increased markedly for recruitment variance. The uncertainty on SSB and Fbar-8 in this year's assessment is higher for the earlier years (before 2015), but has reduced for the most recent estimates (Figure 8.10.3).

## Short term forecast

The intermediate year catch assumption for 2020 used for the short-term forecast in the advice given last year (sum of 2020 TAC of 1090879 tonnes) was slightly lower than the actual 2020 catch reported for WGIWIDE 2021 and used in the present assessment (text table below). The new assessment produced an estimate of the SSB in 2020 which was $7 \%$ higher than the 2020 WGWIDE forecast prediction. This discrepancy in the SSB is explained by the revision of the perception of the abundance at age 6 to $12+$ (Figure 8.10.4) and possibly also by the actual 2020 catch being lower than the value assumed last year. The fishing mortality Fbart-8 for 2020 estimated at the WGWIDE 2020 is 21.9 \% lower than the value estimated by the short-term forecast in the previous assessment also due to the combination of the stock being actually larger than forecasted, and the stock being revised upwards in 2020 (Figure 8.10.1).

|  | Catch (2020) | SSB (2020) | Fbar4-8 (2020) <br> 2020 WGWIDE forecast 1090879 t |
| :--- | :--- | :--- | :--- |
| 2021 WGWIDE assessment | 1039863 t | 391413 t | 0.32 |
| \% difference | $-4.7 \%$ | $7.0 \%$ | 0.25 |

### 8.11 Management Considerations

Details and discussion on quality issues in this year's assessment is given in Section 8.7 above.
From 2001 to 2007, the internationally agreed TACs covered most of the distribution area of the Northeast Atlantic mackerel. From 2008 to 2014, no agreement was reached among the Coastal States on the sharing of the mackerel quotas. In 2014, three of the Coastal States (EU, NO and FO) agreed on a Management Strategy for 2014 to 2018. In November 2018, the agreement from 2014 was extended for two more years until 2020. No agreement on the share of the stock has been reached after Brexit for 2021. Despite various agreements, the total declared quotas in each of the years 2015 to 2020 all exceed the TAC advised by ICES (Figure 8.11.1).

The mackerel in the Northeast Atlantic is traditionally characterised as three distinct 'spawning components': the southern component, the western component and the North Sea component. The basis for the components is derived from tagging experiments (ICES, 1974). However, the methods normally used to identify stocks or components (e.g., ectoparasite infections, blood phenotypes, otolith shapes and genetics) have not been able to demonstrate significant differences between animals from different components. The mackerel in the Northeast Atlantic appears on one hand to mix extensively whilst, on the other hand, exhibit some tendency for homing (Jansen et al., 2013; Jansen and Gislason, 2013). Consequently, it cannot be considered either a panmictic population, nor a population that is composed of isolated components (Jansen and Gislason, 2013). A review of the mackerel in the North Sea, carried out during WKWIDE 2017 (ICES, 2017) concluded that Northeast Atlantic mackerel should be considered as a single population (stock) with individuals that show stronger or weaker affinity for spawning in certain parts of the spawning area.

Nevertheless, stock components are still being used to identify the different spawning areas where mackerel are known to spawn. The trends in the different components is derived from the triennial egg survey in the western and southern area and a dedicated egg survey in the North Sea the year following the western survey.

Since the mid-1970s, ICES has continuously recommended conservation measures for the North Sea component of the Northeast Atlantic mackerel stock (e.g., ICES, 1974; ICES, 1981). The measures advised by ICES to protect the North Sea spawning component (i.e., closed areas and minimum landing size) aimed to promote the conditions that make a recovery of this component possible.

The recommended closure of Division 4.a for fishing during the first half of the year is based on the perception that the western mackerel enter the North Sea in July/August, and remain there until December before migrating to their spawning areas. Updated observations from the late 1990s suggested that this return migration actually started in mid- to late February (Jansen et al., 2012). The EU TAC regulations stated that within the limits of the quota for the western component (ICES Subareas and Divisions 6, 7, 8.a,b,d,e, 5.b (EU), 2.a (non-EU), 12, 14), a certain quantity of this stock may be caught in 4.a between 1 September and 15 February. Up to 2010, $30 \%$ of the EU TAC of mackerel (MAC/2CX14-) could be taken in 4.a. From 2011 until 2014, this percentage increased to $40 \%$ and from 2015 onwards this increased to $60 \%$.

The minimum landing size (MLS) for mackerel is currently set at 30 cm for the North Sea and 20 cm in the western area. The MLS of 30 cm in the North Sea was originally introduced by Norway in 1971 and was intended to protect the very strong 1969 year-class from exploitation in the industrial fishery (Pastoors, 2015). The 30 cm later became the norm for the North Sea MLS while the MLS for mackerel in western waters was set at 20 cm . In the early 1990s, ICES recommended that, because of mixing of juvenile and adult mackerel on western waters fishing grounds, the adoption of a 30 cm minimum landing size for mackerel was not desirable as it could lead to
increased discarding (ICES, 1990; 1991). A substantial part of the catch of (western) NEA mackerel is taken in ICES Division 4.a during the period October until mid-February to which the 30 cm MLS applies even though there is limited understanding on the effectiveness of minimum landing sizes in achieving certain conservation benefits (STECF, 2015).

### 8.12 Ecosystem considerations

An overview of the main ecosystem drivers possibly affecting the different life-stages of Northeast Atlantic mackerel and relevant observations are given in the Stock Annex. The discussion here is limited to recent features of relevance.

## Production (recruitment and growth)

Mackerel recruitment to the fishery (~age 3) was high from year-class 2001, but recently have appeared to be reverting towards a low level. The recruitment index indicates high recruitment at age 0 up to 2020, however, since 2012 the recruitment index has been estimating substantially larger year-classes than what is later estimated at age 3 when they enter the fishery and the other surveys. It is not known if this is a sampling bias or altered mortality of the juveniles between age 0 and 3 .

The increasing stock size was suggested to have an effect through density driven expansion of the spawning area into new areas with Calanus in oceanic areas west of the North European continental shelf (Jansen, 2016). There are several indications of a shift in spawning and mackerel recruitment/larvae and juvenile areas towards northern and north-eastern areas preceding the 2016 mackerel spawning (ICES, 2016; Nøttestad et al., 2018; Bjørdal, 2019). This northerly shift in spawning and recruitment pattern of NEA mackerel seems to have continued also in 2017 (Nøttestad et al., 2018), but spawning in the Norwegian Sea was shown to be of little quantitative significance in 2021 (Burns and O' Hea, WD 15 in Annex 05).

From about 2005 to 2015 mackerel length- and weight-at-age declined substantially for all ages (Jansen and Burns, 2015; Ólafsdóttir et al., 2015). Growth of 0-3 years old mackerel decreased from 1998 to 2012. Mean length at age 0 decreased by 3.6 cm , however the growth differed substantially among cohorts (Jansen and Burns, 2015). For the 3-8 years old mackerel, the average size was reduced by 3.7 cm and 175 g from 2002 to 2013 (Ólafsdóttir et al., 2015). The variations in growth of mackerel in all ages are correlated with mackerel density. Furthermore, the density dependent regulation of growth from younger juveniles to older adult mackerel, appears to reflect the spatial dynamics observed in the migration patterns during the feeding season (Jansen and Burns, 2015; Ólafsdóttir et al., 2015). Growth rates of the juveniles were tightly correlated with the density of juveniles in the nursery areas (Jansen and Burns, 2015). For adult mackerel (age 3-8) growth rates were correlated with the combined effects of mackerel and herring stock sizes (Ólafsdóttir et al., 2015). Conspecific density-dependence was most likely mediated via intensified competition associated with greater mackerel density.

Nevertheless, weight at age of mackerel both from the catches and the surveys have increased during the last few years, particularly for the younger year classes from 1 to 6 years of age (ICES, 2019; 2020).

## Spatial mackerel distribution and timing

In the mid-2000s, the summer feeding distribution of Northeast Atlantic mackerel (Scomber scombrus) in Nordic Seas began expanding into new areas (Nøttestad et al., 2016). During the period 2007-2016 the mackerel distribution range increased three-fold and the centre-of-gravity shifted westward by 1650 km and northward by 400 km . Distribution range peaked in 2014 and was positively correlated to Spawning Stock Biomass (SSB).

After a mackerel stock expansion during the feeding season in summer from 1.3 million $\mathrm{km}^{2}$ in 2007 to at least 2.9 million $\mathrm{km}^{2}$ in 2014, mainly towards western and northern regions of the Nordic seas (Nøttestad et al., 2016), a slight decrease in distribution area of mackerel in the Nordic Seas was observed in 2017 and 2018 with 2.8 million square kilometres (Nøttestad et al., 2017; ICES, 2018a). The mackerel distribution slightly increased to 2.9 million $\mathrm{km}^{2}$ in 2019 (Nøttestad et al., 2019). However, we witnessed a substantial shift in mackerel concentrations and distribution during summers of 2020 and 2021, when no mackerel were registered in Greenland waters, and a substantial decline was documented in Icelandic waters, whereas increased biomasses of mackerel were distributed in the central and northern part of the Norwegian Sea (Nøttestad et al., 2020b), followed by a decrease in 2021 (Nøttestad et al., WD09 in Annex 05). The mackerel was less patchily distributed within the survey area in 2020 compared to 2019. Overall, we have witnessed that mackerel had a much more eastern distribution in 2018 to 2021 compared to 20142017 (ICES, 2018a; Nøttestad et al., 2019; 2020b).

## Spatial mackerel distribution related to environmental conditions

Ólafsdóttir et al. (2018) analysed the IESSNS data from 2007 to 2016 with the following results: Mackerel was present in temperatures ranging from $5^{\circ} \mathrm{C}$ to $15^{\circ} \mathrm{C}$, but preferred areas with temperatures between $9^{\circ} \mathrm{C}$ and $13{ }^{\circ} \mathrm{C}$ according to univariate quotient analysis. Generalized additive models showed that both mackerel occurrence and density were positively related to location, ambient temperature, meso-zooplankton density and SSB, explaining $47 \%$ and $32 \%$ of deviance, respectively. This seem to have changed during 2019 and particularly 2020 where higher concentrations of mackerel were caught in lower temperatures ( $7-8{ }^{\circ} \mathrm{C}$ ) (Nøttestad et al., 2019; 2020b; WD09 in Annex 05). Mackerel relative mean weight-at-length was positively related to location, day-of-year, temperature and SSB, but not with meso-zooplankton density, explaining $40 \%$ of the deviance. Geographical expansion of mackerel during the summer feeding season in Nordic Seas was driven by increasing mackerel stock size and constrained by availability of preferred temperature and abundance of meso-zooplankton. Marine climate with multidecadal variability probably impacted the observed distributional changes but were not evaluated. Our results were limited to the direct effects of temperature, meso-zooplankton abundance, and SSB on distribution range during the last two decades (1997-2016) and should be viewed as such (Olafsdottir et al., 2019). It is not clear what causes this distributional shift, but the SST were 1$2^{\circ} \mathrm{C}$ lower in the western and south-western areas as compared to a 20-years mean (1999-2009), and substantially lower zooplankton concentrations in Icelandic and Greenland waters in 2019 and 2020 than 2018, might partly explain such changes (ICES, 2018a; Nøttestad et al., 2019; 2020a).

## Trophic interactions

There are strong indications for interspecific competition for food between NSS-herring, blue whiting and mackerel (Huse et al., 2012). According to Langøy et al. (2012), Debes et al. (2012), Óskarsson et al. (2015) and Bachiller et al. (2016), the herring may suffer from this competition, as mackerel had higher stomach fullness index than herring and the herring stomach composition is different from previous periods when mackerel stock size was smaller. Langøy et al. (2012) and Debes et al. (2012) also found that mackerel consumed a wider range of prey species than herring. Mackerel may thus be thriving better in periods with low zooplankton abundances. Feeding incidence increased with decreasing temperature as well as stomach filling degree, indicating that feeding activity is highest in areas associated with colder water masses (Bachiller et al., 2016). A bioenergetics model developed by Bachiller et al. (2018) estimated that the NEA mackerel, NSS herring and blue whiting can consume between 122 and 135 million tonnes of zooplankton per year (2005-2010) This is higher than that estimated in previous studies (e.g., Utne et al., 2012; Skjoldal et al., 2004). NEA mackerel feeding rate can consequently be as high as that of the NSS herring in some years. Geographical distribution overlap between mackerel and NSS herring during the summer feeding season is highest in the south-western part of the Norwegian Sea (Faroe and east Icelandic area) (Nøttestad et al., 2016; 2017; Ólafsdóttir et al., 2017). The
spatiotemporal overlap between mackerel and herring was highest in the southern and southwestern part of the Norwegian Sea in 2018 and 2019 (ICES, 2018a, Nøttestad et al., 2019). This is similar as seen in previous years (Nøttestad et al., 2016; 2017). A change was seen in the northern Norwegian Sea in 2019-2021 where we had some increasing overlap between mackerel and herring (mainly 2013- and 2016- year classes) (Nøttestad et al., 2019; 2020; WD09 in Annex 05). There was, on the other hand, practically no overlap between NEA mackerel and NSSH in the central and northern part of the Norwegian Sea in 2018 and previous years, mainly because of very limited amounts of herring in these areas (ICES, 2018a).

There seem to be rather limited spatial overlap between marine mammals and mackerel during summers in the Nordic Seas (Nøttestad et al., 2019; Løviknes, 2019). There is spatial overlap between killer whales and mackerel in the Norwegian Sea, and killer whales are actively hunting for mackerel schools close to the surface during summer (Nøttestad et al., 2014; Nøttestad et al., 2020a). The increase of 0-and 1-groups of NEA mackerel found along major coastlines of Norway both in 2016 and 2017 (Nøttestad et al., 2018) and 2018 (Bjørdal, 2019), has created some interesting new trophic interactions. Increasingly numbers of adult Atlantic bluefin tuna (Thynnus thun$n u s)$, with an average size of approximately 200 kg , have been documented to feed on 0 -group mackerel from the 2016, 2017-year classes during the commercial bluefin tuna fishery in Norway (Boge, 2019; Nøttestad et al., 2020b). Additionally, the new situation of numerous 0-and 1-group mackerel in Norwegian coastal waters in 2018 (Bjørdal, 2019), have created favourable feeding possibilities for larger cod, saithe, marine mammals and seabirds in these waters. Repeated stomach samples from several species document that smaller sized mackerel is now eaten by different predators in northern waters $\left(60-70^{\circ} \mathrm{N}\right)$ (Bjørdal, 2019). Although much fewer 1-groups of NEA mackerel were found along the coast in Norway during the IESSNS 2019 (Nøttestad et al., 2019) and to some extent in 2020 (Nøttestad et al., 2020b) and 2021 (Nøttestad et al., 2021), the Atlantic bluefin tuna is still indeed targeting schools of 1-group mackerel during their intense feeding migration in Norwegian waters (Nøttestad et al., 2020a). The predation pressure and mortality from and increasing Atlantic bluefin tuna stock on NEA mackerel (both juveniles and adults) are unknown, but could have ecological impact on both regional and population level (ICCAT, 2019; Nøttestad et al., 2020b).

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### 8.14 Tables

Table 8.2.4.1. Overview of major existing regulations on mackerel catches.

| Technical measure | National/International level | Specification | Note |
| :---: | :---: | :---: | :---: |
| Catch limitation | Coastal States/NEAFC | 2010-2020 | Not agreed |
| Management strategy (EU, NO, FO agreement London 12. Oct. 2014) | European (EU, NO, FO) | If $S S B>=3.000 .000 t, F=0.24$ <br> If SSB is less than $3.000 .000 \mathrm{t}, \mathrm{F}=$ $0.24 \text { * SSB/3.000.000 }$ <br> TAC should not be changed more than 20\% <br> A party may transfer up to $10 \%$ of unutilised quota to the next year | Not agreed by all parties |
| Management strategy with updated reference points 2019 (EU, NO, FO agreement London 17. Oct. 2019) | European (EU, NO, FO) | If $S S B>=2.500 .000 t, F=0.23$ <br> If SSB is less than $2.500 .000 \mathrm{t}, \mathrm{F}=$ $0.23 * \text { SSB/2.500.000 }$ <br> TAC should not be changed more than $+25 \%$ or $-20 \%$ <br> A party may transfer up to $10 \%$ of unutilised quota to the next year <br> A party may fish up to $10 \%$ beyond the allocated quota, that have to be deduced from next year's quota. | Not agreed by all parties |
| Minimum size (North Sea) | European (EU, NO) | 30 cm in the North Sea |  |
| Minimum size (all areas except North Sea) | European (EU, NO) | 20 cm in all areas except North Sea | 10\% undersized allowed |
| Minimum size | National (NO) | 30 cm in all areas |  |
| Catch limitation | European (EU, NO) | Within the limits of the quota for the western component ( 6,7 , 8.a-b,d,e, 5.b (EC), 2.a (nonEC), $12,14)$, a certain quantity may be taken from 4.a but only during the periods 1 January to 15 February and 1 October to 31 December. |  |
| Area closure | National (UK) | South-West Mackerel Box off Cornwall | Except where the weight of the mackerel does not exceed $15 \%$ by liveweight of the total quantities of mackerel and other marine organisms onboard which have been caught in this area |
| Area limitations | National (IS) | Pelagic trawl fishery only allowed outside of 200 m depth contours around Iceland and/or 12 nm from the coast. |  |


| Technical measure | National/International level | Specification | Note |
| :---: | :---: | :---: | :---: |
| National catch limitations by gear, semester and area | National (ES) | 28.74 \% of the Spanish national quota is assigned for the trawl fishery, $34.29 \%$ for purse seiners and $36.97 \%$ for the artisanal fishery | Since 2015, the trawl fishery has the individual quotas assigned by vessel. |
| Discard prohibition | National (NO, IS, FO) | All discarding is prohibited for Norwegian, Icelandic and Faroese vessels |  |
| Landing Obligation | European | From 2015 onwards a landing obligation for European Union fisheries is in place for small pelagics including mackerel, horse mackerel, blue whiting and herring. <br> In 2016 it was extended to certain demersal fisheries and since 2019 it applies to all TAC species. | There are de minimis exemptions for mackerel caught in bottom-trawl fisheries in the North Western Waters (EC 2018/2034) and in the North Sea (EC 2018/2035). |

Table 8.4.1.1. NE Atlantic Mackerel. ICES estimated catches by area ( $\mathbf{t}$ ). Discards not estimated prior to 1978 (data submitted by Working Group members).

| Year | Subarea 6 |  |  | Subarea 7 and Divisions 8.abde |  |  | Subareas 3 and 4 |  |  | Subareas 125 and 14 |  | Divisions 8.c and 9.a |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg Disc | Catch | Ldg Disc | Catch | Ldg | Disc | Catch |
| 1969 | 4800 |  | 4800 | 47404 |  | 47404 | 739175 |  | 739175 | 7 | 7 | 42526 | 42526 | 833912 |  | 833912 |
| 1970 | 3900 |  | 3900 | 72822 |  | 72822 | 322451 |  | 322451 | 163 | 163 | 70172 | 70172 | 469508 |  | 469508 |
| 1971 | 10200 |  | 10200 | 89745 |  | 89745 | 243673 |  | 243673 | 358 | 358 | 32942 | 32942 | 376918 |  | 376918 |
| 1972 | 13000 |  | 13000 | 130280 |  | 130280 | 188599 |  | 188599 | 88 | 88 | 29262 | 29262 | 361229 |  | 361229 |
| 1973 | 52200 |  | 52200 | 144807 |  | 144807 | 326519 |  | 326519 | 21600 | 21600 | 25967 | 25967 | 571093 |  | 571093 |
| 1974 | 64100 |  | 64100 | 207665 |  | 207665 | 298391 |  | 298391 | 6800 | 6800 | 30630 | 30630 | 607586 |  | 607586 |
| 1975 | 64800 |  | 64800 | 395995 |  | 395995 | 263062 |  | 263062 | 34700 | 34700 | 25457 | 25457 | 784014 |  | 784014 |
| 1976 | 67800 |  | 67800 | 420920 |  | 420920 | 305709 |  | 305709 | 10500 | 10500 | 23306 | 23306 | 828235 |  | 828235 |
| 1977 | 74800 |  | 74800 | 259100 |  | 259100 | 259531 |  | 259531 | 1400 | 1400 | 25416 | 25416 | 620247 |  | 620247 |
| 1978 | 151700 | 15100 | 166800 | 355500 | 35500 | 391000 | 148817 |  | 148817 | 4200 | 4200 | 25909 | 25909 | 686126 | 50600 | 736726 |
| 1979 | 203300 | 20300 | 223600 | 398000 | 39800 | 437800 | 152323 | 500 | 152823 | 7000 | 7000 | 21932 | 21932 | 782555 | 60600 | 843155 |
| 1980 | 218700 | 6000 | 224700 | 386100 | 15600 | 401700 | 87931 |  | 87931 | 8300 | 8300 | 12280 | 12280 | 713311 | 21600 | 734911 |
| 1981 | 335100 | 2500 | 337600 | 274300 | 39800 | 314100 | 64172 | 3216 | 67388 | 18700 | 18700 | 16688 | 16688 | 708960 | 45516 | 754476 |
| 1982 | 340400 | 4100 | 344500 | 257800 | 20800 | 278600 | 35033 | 450 | 35483 | 37600 | 37600 | 21076 | 21076 | 691909 | 25350 | 717259 |
| 1983 | 320500 | 2300 | 322800 | 235000 | 9000 | 244000 | 40889 | 96 | 40985 | 49000 | 49000 | 14853 | 14853 | 660242 | 11396 | 671638 |
| 1984 | 306100 | 1600 | 307700 | 161400 | 10500 | 171900 | 43696 | 202 | 43898 | 98222 | 98222 | 20208 | 20208 | 629626 | 12302 | 641928 |


| Year | Subarea 6 |  |  | Subarea 7 and Divisions 8.abde |  |  | Subareas 3 and 4 |  |  | Subareas 125 and 14 |  | Divisions 8.c and 9.a |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg Disc | Catch | Ldg Disc | Catch | Ldg | Disc | Catch |
| 1985 | 388140 | 2735 | 390875 | 75043 | 1800 | 76843 | 46790 | 3656 | 50446 | 78000 | 78000 | 18111 | 18111 | 606084 | 8191 | 614275 |
| 1986 | 104100 |  | 104100 | 128499 |  | 128499 | 236309 | 7431 | 243740 | 101000 | 101000 | 24789 | 24789 | 594697 | 7431 | 602128 |
| 1987 | 183700 |  | 183700 | 100300 |  | 100300 | 290829 | 10789 | 301618 | 47000 | 47000 | 22187 | 22187 | 644016 | 10789 | 654805 |
| 1988 | 115600 | 3100 | 118700 | 75600 | 2700 | 78300 | 308550 | 29766 | 338316 | 120404 | 120404 | 24772 | 24772 | 644926 | 35566 | 680492 |
| 1989 | 121300 | 2600 | 123900 | 72900 | 2300 | 75200 | 279410 | 2190 | 281600 | 90488 | 90488 | 18321 | 18321 | 582419 | 7090 | 589509 |
| 1990 | 114800 | 5800 | 120600 | 56300 | 5500 | 61800 | 300800 | 4300 | 305100 | 118700 | 118700 | 21311 | 21311 | 611911 | 15600 | 627511 |
| 1991 | 109500 | 10700 | 120200 | 50500 | 12800 | 63300 | 358700 | 7200 | 365900 | 97800 | 97800 | 20683 | 20683 | 637183 | 30700 | 667883 |
| 1992 | 141906 | 9620 | 151526 | 72153 | 12400 | 84553 | 364184 | 2980 | 367164 | 139062 | 139062 | 18046 | 18046 | 735351 | 25000 | 760351 |
| 1993 | 133497 | 2670 | 136167 | 99828 | 12790 | 112618 | 387838 | 2720 | 390558 | 165973 | 165973 | 19720 | 19720 | 806856 | 18180 | 825036 |
| 1994 | 134338 | 1390 | 135728 | 113088 | 2830 | 115918 | 471247 | 1150 | 472397 | 72309 | 72309 | 25043 | 25043 | 816025 | 5370 | 821395 |
| 1995 | 145626 | 74 | 145700 | 117883 | 6917 | 124800 | 321474 | 730 | 322204 | 135496 | 135496 | 27600 | 27600 | 748079 | 7721 | 755800 |
| 1996 | 129895 | 255 | 130150 | 73351 | 9773 | 83124 | 211451 | 1387 | 212838 | 103376 | 103376 | 34123 | 34123 | 552196 | 11415 | 563611 |
| 1997 | 65044 | 2240 | 67284 | 114719 | 13817 | 128536 | 226680 | 2807 | 229487 | 103598 | 103598 | 40708 | 40708 | 550749 | 18864 | 569613 |
| 1998 | 110141 | 71 | 110212 | 105181 | 3206 | 108387 | 264947 | 4735 | 269682 | 134219 | 134219 | 44164 | 44164 | 658652 | 8012 | 666664 |
| 1999 | 116362 |  | 116362 | 94290 |  | 94290 | 313014 |  | 313014 | 72848 | 72848 | 43796 | 43796 | 640311 |  | 640311 |
| 2000 | 187595 | 1 | 187595 | 115566 | 1918 | 117484 | 285567 | 165 | 304898 | 92557 | 92557 | 36074 | 36074 | 736524 | 2084 | 738608 |
| 2001 | 143142 | 83 | 143142 | 142890 | 1081 | 143971 | 327200 | 24 | 339971 | 67097 | 67097 | 43198 | 43198 | 736274 | 1188 | 737462 |


| Year | Subarea 6 |  |  | Subarea 7 and Divisions 8.abde |  |  | Subareas 3 and 4 |  |  | Subareas 125 and 14 |  |  | Divisions 8.c and 9.a |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch |
| 2002 | 136847 | 12931 | 149778 | 102484 | 2260 | 104744 | 375708 | 8583 | 394878 | 73929 |  | 73929 | 49576 |  | 49576 | 749131 | 23774 | 772905 |
| 2003 | 135690 | 1399 | 137089 | 90356 | 5712 | 96068 | 354109 | 11785 | 365894 | 53883 |  | 53883 | 25823 | 531 | 26354 | 659831 | 19427 | 679288 |
| 2004 | 134033 | 1705 | 134738 | 103703 | 5991 | 109694 | 306040 | 11329 | 317369 | 62913 | 9 | 62922 | 34840 | 928 | 35769 | 640529 | 19962 | 660491 |
| 2005 | 79960 | 8201 | 88162 | 90278 | 12158 | 102436 | 249741 | 4633 | 254374 | 54129 |  | 54129 | 49618 | 796 | 50414 | 523726 | 25788 | 549514 |
| 2006 | 88077 | 6081 | 94158 | 66209 | 8642 | 74851 | 200929 | 8263 | 209192 | 46716 |  | 46716 | 52751 | 3607 | 56358 | 454587 | 26594 | 481181 |
| 2007 | 110788 | 2450 | 113238 | 71235 | 7727 | 78962 | 253013 | 4195 | 257208 | 72891 |  | 72891 | 62834 | 1072 | 63906 | 570762 | 15444 | 586206 |
| 2008 | 76358 | 21889 | 98247 | 73954 | 5462 | 79416 | 227252 | 8862 | 236113 | 148669 | 112 | 148781 | 59859 | 750 | 60609 | 586090 | 37075 | 623165 |
| 2009 | 135468 | 3927 | 139395 | 88287 | 2921 | 91208 | 226928 | 8120 | 235049 | 163604 |  | 163604 | 107747 | 966 | 108713 | 722035 | 15934 | 737969 |
| 2010 | 106732 | 2904 | 109636 | 104128 | 4614 | 108741 | 246818 | 883 | 247700 | 355725 | 5 | 355729 | 50826 | 4640 | 55466 | 864229 | 13045 | 877272 |
| 2011 | 160756 | 1836 | 162592 | 51098 | 5317 | 56415 | 301746 | 1906 | 303652 | 398132 | 28 | 398160 | 26337 | 1807 | 28144 | 938070 | 10894 | 948963 |
| 2012 | 121115 | 952 | 122067 | 65728 | 9701 | 75429 | 218400 | 1089 | 219489 | 449325 | 1 | 449326 | 29809 | 3431 | 33240 | 884377 | 15174 | 899551 |
| 2013 | 132062 | 273 | 132335 | 49871 | 1652 | 51523 | 260921 | 337 | 261258 | 465846 | 15 | 465861 | 24867 | 2455 | 27322 | 933567 | 4732 | 938299 |
| 2014 | 180068 | 340 | 180408 | 93709 | 1402 | 95111 | 383887 | 334 | 384221 | 684082 | 91 | 684173 | 53591 | 4284 | 57875 | 1395337 | 6451 | 1401788 |
| 2015 | 134728 | 30 | 134757 | 98563 | 3155 | 101718 | 295877 | 34 | 295911 | 632493 | 78 | 632571 | 43735 | 7133 | 50869 | 1205396 | 10431 | 1215827 |
| 2016 | 206326 | 200 | 206526 | 37300 | 1927 | 39227 | 248041 | 570 | 248611 | 563440 | 54 | 563494 | 39056 | 3220 | 42276 | 1094163 | 5971 | 1100135 |
| 2017 | 225959 | 151 | 226110 | 21128 | 1992 | 23119 | 269404 | 400 | 269804 | 603806 | 62 | 603869 | 36512 | 227 | 36739 | 1156809 | 2832 | 1159641 |
| 2018 | 157239 | 90 | 157329 | 32037 | 1611 | 33649 | 341527 | 620 | 342147 | 455689 | 51 | 455740 | 33761 | 518 | 34279 | 1020254 | 2890 | 1023144 |
| 2019 | 122995 | 144 | 123139 | 32840 | 5902 | 38742 | 307235 | 812 | 308047 | 345019 | 18 | 345037 | 23832 | 931 | 24763 | 831920 | 7807 | 839727 |
| 2020 | 130577 | 341 | 130918 | 48806 | 8065 | 56871 | 456479 | 732 | 457211 | 356985 |  | 356985 | 37386 | 143 | 37529 | 1030232 | 9280 | 1039513 |

Table 8.4.2.1. NE Atlantic Mackerel. ICES estimated catch ( $\mathbf{t}$ ) in Subareas 1, 2, 5 and 14, 2000-2020 (Data submitted by Working Group members).

| YearDen- <br> mark | Esto- <br> nia | Faroe <br> Islands | France |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| Year | Denmark | Esto- <br> nia | Faroe Islands | France | Germany | Greenland | Iceland | Ire- <br> land | Lithuania | Netherlands | Norway | Po- <br> land | Swe- <br> den | United Kingdom | Russia | Mis-reported | Unallocated | Discards | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | 2 |  | 66194 |  | 4064 | 46388 | 167366 |  |  | 7671 | 167739 |  | 1720 | 4601 | 138061 |  |  | 62 | 603869 |
| 2018 | 289 |  | 52061 | 733 | 577 | 62973 | 168330 |  |  | 2697 | 46853 | 2 | 910 | 2009 | 118255 |  |  | 51 | 455740 |
| 2019 |  |  | 37418 |  | 190 | 30241 | 128008 |  |  | 13 | 22605 |  |  |  | 126543 |  |  | 18 | 345036 |
| 2020 |  |  | 33291 | 8 | 206 | 26555 | 151534 |  | 2 | 0.73 | 15937 | 0.044 | 220 | 426 | 128805 |  |  | 0.05 | 356985 |

Table 8.4.2.2. NE Atlantic Mackerel. ICES estimated catch ( t ) in the North Sea, Skagerrak and Kattegat (Subarea 4 and Division 3.a), 2000-2020 (Data submitted by Working Group members).

| Year | Belgium | Den- <br> mark | Faroe Islands | France | Germany. | Ire- <br> land | Lithuania | Netherlands | Norway | Po- <br> land | Sweden | United Kingdom | Russia | Misreported (Area 6.a) | Unal-located | Discards | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 146 | 27720 | 10614 | 1588 | 78 | 9956 |  | 2262 | 142320 |  | 49941 | 58282 | 1672 | 8591 | 34761 | 1912 | 304896 |
| 2001 | 97 | 21680 | 18751 | 1981 | 4514 | 10284 |  | 2441 | 158401 |  | 5090 | 52988 | 1 | 39024 | 24873 | 24 | 339970 |
| 2002 | 22 | 343751 | 12548 | 2152 | 3902 | 20715 |  | 11044 | 161621 |  | 52321 | 61781 |  | 49918 | 22985 | 8583 | 394878 |
| 2003 | 2 | 275081 | 11754 | 1467 | 4859 | 17145 |  | 6784 | 150858 |  | 4450 | 67083 |  | 62928 | -730 | 11785 | 365894 |
| 2004 | 4 | 25665 | 11705 | 1538 | 4515 | 18901 |  | 6366 | 147068 |  | 4437 | 62932 |  | 23692 | -783 | 11329 | 317369 |
| 2005 | 1 | 232121 | 9739 | 1004 | 4442 | 15605 |  | 3915 | 106434 | 109 | 3204 | 37118 | 4 | 37911 | 7043 | 4633 | 254374 |
| 2006 | 3 | 242191 | 12008 | 285 | 2389 | 4125 |  | 4093 | 113079 |  | 3209 | 28628 |  | 8719 | 171 | 8263 | 209192 |
| 2007 | 1 | 252171 | 11818 | 7549 | 5383 | 13337 |  | 5973 | 131191 |  | 38581 | 46264 |  |  | 2421 | 4195 | 257208 |
| 2008 | 2 | 26716 | 7627 | 490 | 4668 | 11628 |  | 1980 | 114102 |  | 36641 | 37055 |  | 17280 | 2039 | 8862 | 236111 |
| 2009 | 3 | 23491 | 6648 | 1493 | 5158 | 12901 |  | 2039 | 118070 |  | 73031 | 47863 |  | 1959 | -629 | 8120 | 235049 |

$\left.\begin{array}{llllllllllll}\hline \text { Year } & \text { Belgium } & \begin{array}{l}\text { Den- } \\ \text { mark }\end{array} & \begin{array}{l}\text { Faroe Is- } \\ \text { lands }\end{array} & \text { France } & \begin{array}{l}\text { Ger- } \\ \text { many. }\end{array} & \begin{array}{l}\text { lre- } \\ \text { land }\end{array} & \begin{array}{l}\text { Lithua- } \\ \text { nia }\end{array} & \begin{array}{l}\text { Nether- } \\ \text { lands }\end{array} & \text { Norway } & \begin{array}{l}\text { Po- } \\ \text { land }\end{array} & \begin{array}{l}\text { Sweden }\end{array} \begin{array}{l}\text { United } \\ \text { King- } \\ \text { dom }\end{array} \\ \hline 2010 & 27 & 36552 & 4639 & 686 & 25621 & 14639 & 1300 & 129064 & \begin{array}{l}\text { Misre- } \\ \text { ported } \\ \text { (Area } \\ \text { 6.a) }\end{array} \\ \text { lo- } \\ \text { cated }\end{array}\right\}$

Table 8.4.2.3. NE Atlantic Mackerel. ICES estimated catch ( t ) in the Western area (Subareas 6 and 7 and Divisions 8.a,b,d,e), 2000 - 2020 (Data submitted by Working Group members).

| Year | Belgium | Denmark | Faroe Islands | France | Ger- <br> many | Greenland | Ice- <br> land | Ire- <br> land | Lithuania | Neth-er- | Norway | Poland | Portugal | Rus- <br> sia | Spain | Swe- <br> den | United Kingdom | Misreported | Unal-located | Discards | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 |  | 82 | 4863 | 17857 | 22901 |  |  | 61277 |  | 30123 |  |  |  |  | 4500 |  | 126620 | -3775 | 31564 | 1920 | 297932 |
| 2001 |  | 835 | 2161 | 18975 | 20793 |  |  | 60168 |  | 33654 |  |  |  |  | 4063 |  | 139589 | $39024$ | 37952 | 1164 | 280553 |


| YearBel- <br> gium | Den- <br> mark | Faroe <br> Is- <br> lands | France |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| Year | Belgium | Denmark | Faroe Is- <br> lands | France | Germany | Greenland | Ice- <br> land | Ireland | Lithuania | Neth-er- <br> lands | Norway | Po- <br> land | Portugal | Rus- <br> sia | Spain | Sweden | United Kingdom | Misreported | Unal-located | Discards | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | 21 | 12569 | 20559 | 16925 | 9608 |  | 48957 | 2 | 18694 | 2657 |  |  |  |  | 786 |  | 116308 |  |  | 2142 | 249229 |
| 2018 | 58 | 8194 | 13543 | 13974 | 7214 |  |  | 42181 |  | 13851 | 4639 | 14 |  |  | 1269 |  | 84327 |  | 13 | 1701 | 190978 |
| 2019 | 53 | 5189 | 7787 | 12371 | 8936 |  | 69 | 51635 |  | 13727 | 1420 | 2312 | 46 | 1 | 1217 | 805 | 50267 |  |  | 6046 | 161879 |
| 2020 | 49 | 4110 | 2913 | 12816 | 8878 | 22 |  | 58720 |  | 11895 | 221 | 5286 | 35 | 10 | 1784 |  | 72645 |  |  | 8405 | 187788 |

Table 8.4.2.4. NE Atlantic Mackerel. ICES estimated catch ( $\mathbf{t}$ ) in Divisions 8.c and 9.a, 2000-2020 (Data submitted by Working Group members). 9.b is included in 2020.

| Country | France 8.c | Portugal 9.a | Portugal 8.c | Russia 9.b | Spain 8.c | Spain 9.a | Discards 8.c | Discards 9.a | Unallocated 8.c | Unallocated 9.a | Total 9.a | Total 8c and 9a |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 |  | 2253 |  |  | 30061 | 3760 |  | 6013 |  |  | 12026 | 42087 |
| 2001 |  | 3119 |  |  | 38205 | 1874 |  |  |  |  | 4993 | 43198 |
| 2002 |  | 2934 |  |  | 38703 | 7938 |  |  |  |  | 10873 | 49575 |
| 2003 | 226 | 2749 |  |  | 17384 | 5464 | 531 |  |  |  | 8213 | 26354 |
| 2004 | 177 |  | 2289 |  |  |  | 928 |  | 28429 | 3946 | 6234 | 35768 |
| 2005 | 151 |  | 1509 |  |  |  | 391 | 405 | 42851 | 5107 | 7021 | 50414 |
| 2006 | 43 |  | 2620 |  | 43063 | 7025 | 3606 | 1 |  |  | 9646 | 56358 |
| 2007 | 55 |  | 2605 |  | 53401 | 6773 | 156 | 916 |  |  | 10293 | 63906 |
| 2008 | 168 |  | 2381 |  | 50455 | 6855 | 73 | 677 |  |  | 9913 | 60609 |
| 2009 | 383 |  | 1753 |  | 91043 | 14569 | 725 | 241 |  |  | 16562 | 108713 |
| 2010 | 392 | 1758 | 2363 |  | 38858 | 7347 | 4408 | 232 |  | 108 | 10049 | 55466 |
| 2011 | 44 | 2302 | 962 |  | 14709 | 2759 | 563 | 1245 | 4691 | 871 | 5836 | 28146 |
| 2012 | 283 | 4868 | 824 |  | 17768 | 845 | 2187 | 1244 | 4144 | 1076 | 3989 | 33239 |
| 2013 | 220 | 5134 | 254 |  | 14617 | 1162 | 1428 | 1027 | -573 | 4053 | 6497 | 27322 |
| 2014 | 171 | 7334 | 618 |  | 33783 | 2227 | 2821 | 1463 | 8795 | 662 | 4308 | 57874 |
| 2015 | 21 | 6836 | 1456 |  | 29726 | 3853 | 4724 | 2409 | 11 | 1831 | 9550 | 50867 |
| 2016 | 106 | 6069 | 619 |  | 26553 | 2229 | 2469 | 751 | 1357 | 2123 | 5722 | 42276 |
| 2017 | 83 | 3697 | 634 |  | 30893 | 1206 | 84 | 143 |  |  | 1983 | 36740 |


| Country | France 8.c | Portugal 9.a | Portugal 8.c | Russia 9.b | Spain 8.c | Spain 9.a | Discards 8.c | Discards 9.a | Unallocated 8.c | Unallocated 9.a | Total 9.a | Total 8c and 9a |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2018 | 50 | 3709 | 855 |  | 27190 | 1656 | 324 | 194 | 300 | 2736 | 34279 |  |
| 2019 | 43 | 3188 | 706 |  | 19148 | 747 | 760 | 172 | 1625 | 24764 |  |  |
| 2020 | 96 | 4189 | 575 | 3 | 31143 | 1379 | 28 | 115 | 2069 | 37529 |  |  |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2020 (Q1-Q4).

| Age | 1 | 2.a | 2.12 | 2.a2 | 3.a | 3.b | 3.c | 3.d | 4.a | 4.b | 4.c | 5.a |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  | 249 |  |  |  | 244 | 217 | 197 |  |
| 1 |  | 280 | 263 | 263 | 292 | 289 | 292 | 292 | 296 | 295 | 295 |  |
| 2 | 335 | 327 | 329 | 329 | 321 | 324 | 317 | 323 | 320 | 321 | 322 |  |
| 3 | 348 | 331 | 331 | 331 | 330 | 336 | 327 | 320 | 332 | 323 | 326 | 353 |
| 4 | 358 | 341 | 343 | 343 | 343 | 348 | 340 | 329 | 344 | 338 | 341 | 351 |
| 5 | 353 | 345 | 357 | 357 | 354 | 360 | 355 | 348 | 356 | 350 | 354 | 367 |
| 6 | 371 | 360 | 368 | 368 | 363 | 368 | 363 | 366 | 364 | 351 | 357 | 369 |
| 7 | 373 | 364 | 365 | 366 | 372 | 375 | 372 | 381 | 371 | 365 | 370 | 373 |
| 8 | 379 | 369 | 371 | 371 | 376 | 378 | 376 | 384 | 375 | 366 | 376 | 376 |
| 9 | 385 | 374 | 377 | 377 | 378 | 380 | 379 | 389 | 378 | 372 | 374 | 377 |
| 10 | 390 | 373 | 374 | 374 | 384 | 389 | 386 | 386 | 383 | 382 | 383 | 379 |
| 11 |  | 377 | 376 | 376 | 384 | 391 | 389 | 397 | 388 | 383 | 384 | 384 |
| 12 |  | 382 | 389 | 389 | 391 | 396 | 399 | 390 | 391 | 389 | 380 | 390 |
| 13 |  | 385 | 380 | 381 | 395 | 399 | 399 | 403 | 393 | 390 | 391 | 392 |
| 14 |  | 390 | 392 | 392 | 396 | 402 | 415 | 393 | 397 | 390 | 392 | 394 |
| 15+ |  | 398 | 395 | 395 | 403 | 406 | 406 | 402 | 397 | 396 | 402 | 390 |


| Age | 5.b | 5.b. 1 | 6.a | 6.b | 7.a | 7.b | 7.c | 7.d | 7.9 | 7.h | 7.j | 7.k |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  | 173 |  |  |  |  |
| 1 |  |  | 174 | 248 |  | 295 |  | 283 |  |  |  |  |
| 2 |  |  | 296 | 314 |  | 304 | 306 | 318 |  |  |  |  |
| 3 | 353 | 353 | 328 | 325 |  | 328 | 325 | 330 | 113 | 174 | 335 | 345 |
| 4 | 352 | 351 | 342 | 344 | 131 | 341 | 339 | 343 | 268 | 287 | 336 | 358 |
| 5 | 359 | 364 | 359 | 357 | 306 | 361 | 347 | 359 | 361 | 361 | 365 | 365 |
| 6 | 367 | 368 | 365 | 365 | 353 | 367 | 365 | 371 | 313 | 306 | 369 | 369 |
| 7 | 369 | 371 | 372 | 372 | 362 | 373 | 376 | 372 | 352 | 361 | 370 | 380 |
| 8 | 371 | 374 | 376 | 375 | 350 | 375 | 397 | 383 | 362 | 369 | 380 | 381 |
| 9 | 372 | 375 | 377 | 378 | 381 | 376 | 382 | 379 | 379 | 379 | 379 | 379 |


| Age | 5.b | 5.b.1 | 6.a | 6.b | 7.a | 7.b | 7.c | 7.d | 7.g | 7.h | 7.j | 7.k |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10 | 374 | 377 | 383 | 385 | 388 | 382 | 393 | 385 | 387 | 385 | 384 | 398 |
| 11 | 374 | 380 | 391 | 393 | 402 | 387 | 403 | 424 | 409 | 433 | 423 | 399 |
| 12 | 385 | 388 | 394 | 396 | 373 | 387 | 387 | 405 | 399 | 403 | 402 | 395 |
| 13 | 389 | 391 | 397 | 399 |  | 389 | 392 | 393 | 395 | 395 |  |  |
| 14 | 391 | 393 | 404 | 413 |  | 388 | 388 | 396 | 425 | 425 | 425 | 425 |
| $15+$ | 380 | 388 | 409 | 412 |  | 401 | 401 | 416 |  |  |  |  |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2020 (Q1-Q4) continued.

| Age | $8 . \mathrm{a}$ | 8.b | 8.c | 8.c.E | 8.c.W | 8.d | $8 . e$ | 9.1 | 9.a.N | 12.c | All |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 177 | 177 | 202 | 186 | 0 |  |  | 266 | 194 |  | 192 |
| 1 | 287 | 246 | 252 | 297 | 322 | 322 | 307 | 288 | 251 |  | 125 |
| 2 | 305 | 295 | 290 | 308 | 323 | 322 | 295 | 294 | 280 |  | 291 |
| 3 | 331 | 335 | 321 | 338 | 335 | 337 | 316 | 325 | 306 | 335 | 320 |
| 4 | 357 | 353 | 343 | 353 | 353 | 339 | 351 | 354 | 350 | 333 | 342 |
| 5 | 361 | 351 | 354 | 364 | 370 | 357 | 369 | 371 | 357 | 365 | 351 |
| 6 | 362 | 361 | 368 | 366 | 380 | 363 | 378 | 377 | 361 | 368 | 363 |
| 7 | 358 | 362 | 374 | 374 | 384 | 366 | 381 | 385 | 373 | 369 | 369 |
| 8 | 379 | 377 | 377 | 378 | 384 | 379 | 382 | 377 | 377 | 380 | 374 |
| 9 | 374 | 379 | 382 | 379 | 385 | 375 | 385 | 395 | 379 | 378 | 377 |
| 10 | 374 | 375 | 391 | 387 | 390 | 376 | 389 | 405 | 389 | 382 | 380 |
| 11 | 372 | 374 | 394 | 392 | 415 | 386 | 415 | 405 | 399 | 455 | 384 |
| 12 | 384 | 390 | 403 | 397 | 411 | 391 | 415 |  | 401 | 405 | 388 |
| 13 | 382 | 382 | 400 | 425 |  | 382 |  | 420 |  |  | 390 |
| 14 | 396 | 396 | 410 | 435 |  | 396 |  |  |  |  | 393 |
| 15+ | 405 | 405 | 432 | 432 |  | 405 |  |  | 420 |  | 398 |

Table 8.6.1.1.1. International mackerel and horse mackerel egg survey in the western and southern areas: Periods and area assignments for countries/institutes by week for the $\mathbf{2 0 2 2}$ survey. Area assignments and dates are provisional.

| Week | Starts | Area <br> 9a | Cantabrian Sea | Biscay | Celtic <br> sea | West of Ireland | West of Scotland | Northern area | Period |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 09-Jan-22 |  |  |  |  |  |  |  | 1 |
| 4 | 16-Jan-22 | PO1 |  |  |  |  |  |  | 2 |
| 5 | 23-Jan-22 | PO1 |  |  |  |  |  |  | 2 |
| 6 | 30-Jan-22 | PO1 |  |  |  |  |  |  | 2 |
| 7 | 06-Feb-22 | PO1 |  |  |  |  |  |  | 2 |
| 8 | 13-Feb-22 | PO1 |  |  |  |  |  |  | 2 |
| 9 | 20-Feb-22 | PO1 |  |  |  | SCO (IBTS) | $\begin{aligned} & \text { SCO } \\ & \text { (IBTS) } \end{aligned}$ |  | 2 |
| 10 | 27-Feb-22 |  |  |  |  | SCO (IBTS) | $\begin{aligned} & \text { SCO } \\ & \text { (IBTS) } \end{aligned}$ |  | 2 |
| 11 | 06-Mar-22 |  |  | IEO1 | IRL 1 | IRL 1 | IRL 1 |  | 3 |
| 12 | 13-Mar-22 |  |  | IEO1 | IRL 1 | IRL 1 | IRL 1 |  | 3 |
| 13 | 20-Mar-22 |  | IEO1 | AZTI1 | GER1 | IRL 1 | IRL 1 |  | 3 |
| 14 | 27-Mar-22 |  | IEO1 | AZTI 1 | GER1 | GER1 |  |  | 3 |
| 15 | 03-Apr-22 |  |  | AZTI1 | GER1 | GER1 |  |  | 3 |
| 16 | 10-Apr-22 |  | IEO2 | IEO2 | GER2 | GER 2 /SCO1 | SCO1 |  | 4 |
| 17 | 17-Apr-22 |  | IEO2 | IEO2 | GER2 | GER 2 /SCO1 | SCO1 |  | 4 |
| 18 | 24-Apr-22 |  | IEO2 | IEO2 | GER2 | GER 2 /SCO1 | SCO1 |  | 4 |
| 19 | 1-May-22 |  | IEO2/AZTI2 <br> (DEPM) | IEO2 |  |  |  |  | 4 |
| 20 | 8-May-22 |  | AZTI2 <br> (DEPM) | AZTI2 (DEPM)/ NED1 | NED1 | NED1 / SCO2 | SCO2 | NOR | 5 |
| 21 | 15-May-22 |  |  | AZTI2 (DEPM)/ <br> NED1 | NED1 | NED1 / SCO2 | SCO2 | NOR | 5 |
| 22 | 22-May-22 |  |  | AZTI2 (DEPM)/ <br> NED1 | NED1 | NED1 / SCO2 | SCO2 | NOR | 5 |
| 23 | 29-May-22 |  |  |  |  |  |  | FAR | 6 |
| 24 | 5-Jun-22 |  |  | NED2 | NED2 | IRL2 | IRL2 | FAR | 6 |
| 25 | 12-Jun-22 |  |  | NED2 | NED2 | IRL2 | IRL2 | FAR | 6 |
| 26 | 19-Jun-22 |  |  | NED2 | NED2 | IRL2 | IRL2 |  | 6 |
| 27 | 26-Jun-22 |  |  |  |  |  |  |  | 6 |
| 28 | 3-Jul-22 |  |  |  | SCO3 | SCO3 | SCO3 |  | 7 |


| Week | Starts | Area <br> 9 a | Cantabrian <br> Sea | Biscay | Celtic <br> sea | West of Ire- <br> land | West of <br> Scotland | Northern <br> area |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 29 | $10-$ Jul-22 |  |  | SCO3 | SCO3 | SCO3 | 7 |  |
| 30 | 17-Jul-22 |  | SCO3 | SCO3 | SCO3 | 7 |  |  |
| 31 | $24-J u l-22$ |  | $S C O 3$ | SCO3 | SCO3 | 6 |  |  |

Table 8.6.1.3.1. Daily egg production estimate (stage 1A) for mackerel in the North Sea using the DEPM.

| Year | DEP ${ }^{10^{13}}$ | CV DEP |
| :--- | :--- | :--- |
| 2021 | 1.28 | $16 \%$ |

Table 8.6.2.1. Model parameter estimates and standard errors.

| Symbol | Description | Unit | Estimate | Std.Error |
| :--- | :--- | :--- | :--- | :--- |
| T | Decorrelation time | year | 1,9 | 0.3 |
| H | Spatial decorrelation distance | km | 455 | 82 |
| $W S$ | Log Wing spread | nmi | -1.0 | 0.6 |
| $\sigma_{N}^{2}$ | Variance of the nugget effect | 1 | 3.7 | 5.3 |
| $\sigma_{x y}^{2}$ | Spatial variance parameter | 1 | 5.4 |  |
|  | (year specific surfaces) | 1 |  |  |

Table 8.7.1.1. NE Atlantic mackerel. Input data and parameters and the model configurations for the assessment.

| Input data types and characteristics: |  |  |  |
| :--- | :--- | :--- | :--- |
| Name | Year range | Age <br> range | Variable from year to year |
| Catch in tonnes | $1980-2020$ | Yes |  |
| Catch-at-age in numbers | $1980-2020$ | $0-12+$ | Yes |
| Weight-at-age in the commercial catch $1980-2020$ $0-12+$ <br> Weight-at-age of the spawning stock <br> at spawning time. $1980-2020$ Yes <br> Proportion of natural mortality before <br> spawning $1980-2021$ $0-12+$ <br> Proportion of fishing mortality before <br> spawning $1980-2021$ Yes <br> Proportion mature-at-age $1980-2021$ $0-12+$ <br> Natural mortality Yes Yes | 1980-2021 Nixed at 0.15 |  |  |


| Type | Name |  | Year range |  | Age range |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Survey (SSB) | ICES Triennial Mackerel and Horse Mackerel Egg Survey |  | 1992, 1995, 1998, 2001, 2004, 2007, 2010, 2013,2016,2019. |  | Not applicable (give SSB) |
| Survey (abundance index) | IBTS Recruitment index (log transformed) |  | 1998-2020 |  | Age 0 |
| Survey <br> (abundance index) | International Ecosystem Summer Survey in the Nordic Seas (IESSNS) |  | 2010, 2012-2021 |  | Ages 3-11 |
| Tagging/recapture | Norwegian tagging program |  | Steal tags : 1980 (release year)2006 (recapture years) <br> RFID tags : 2013 (release year) 2020 (recapture year) |  | Ages 5 and older (age at release) |
| SAM parameter configuration : |  |  |  |  |  |
| Setting |  | Value |  | Description |  |
| Coupling of fishing mortality states |  | 1/2/3/4/5/6/7/8/8/8/8/8/8 |  | Different F states for ages 0 to 6, one same F state for ages 7 and older |  |
| Correlated random walks for the fish ing mortalities |  | 0 |  | F random walk of different ages are independent |  |
| Coupling of catchability parameters |  | 0/0/0/0/0/0/0/0/0/0/0/0/0 <br> 1/0/0/0/0/0/0/0/0/0/0/0/0 <br> 2/0/0/0/0/0/0/0/0/0/0/0/0 <br> 0/0/0/3/4/5/6/7/8/9/10/10/0 |  | No catchability p <br> One catchability the egg <br> One catchability the recruitment <br> One catchability group estimated to11) | eter for the catches meter estimated for meter estimated for meter for each age he IESSNS (age 3 |
| Power law model |  | 0 |  | No power law model used for any of the surveys |  |
| Coupling of fishing mortality random walk variances |  | 1/2/3/3/3/3/3/3/3/3/3/3/3 |  | Separate F random walk variances for age 0 , age 1 and a same variance for older ages |  |
| Coupling of log abundance random walk variances |  | 1/2/2/2/2/2/2/2/2/2/2/2/2 |  | Same variance used for the log abundance random walk of all ages except for the recruits (age 0) |  |
| Coupling of the observation variances |  | 1/2/3/3/3/3/3/3/3/3/3/3/3 <br> 0/0/0/0/0/0/0/0/0/0/0/0/0 <br> 4/0/0/0/0/0/0/0/0/0/0/0/0 <br> 0/0/0/5/6/6/6/6/6/6/6/6/0 |  | Separate observation variances for age 0 and 1 than for the older ages in the catches <br> One observation variance for the egg survey <br> One observation variance for the recruitment index |  |
| Stock recruitment model |  | 0 |  | No stock-recruiment model |  |


| Correlation structure | "ID", "ID", "ID", "AR" |
| :--- | :--- | | Auto-regressive correlation structure for |
| :--- |
| the IESSNS index, independent observa- |
| tions assumed for the other data sources |

## Table 8.7.1.2. NE Atlantic Mackerel. CATCH IN NUMBER

|  | year |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 0 | 33101 | 56682 | 11180 | 7333 | 287287 | 81799 | 49983 | 7403 | 57644 | 65400 |
| 1 | 411327 | 276229 | 213936 | 47914 | 31901 | 268960 | 58126 | 40126 | 152656 | 64263 |
| 2 | 393025 | 502365 | 432867 | 668909 | 86064 | 20893 | 424563 | 156670 | 137635 | 312739 |
| 3 | 64549 | 231814 | 472457 | 433744 | 682491 | 58346 | 38387 | 663378 | 190403 | 207689 |
| 4 | 328206 | 32814 | 184581 | 373262 | 387582 | 445357 | 76545 | 56680 | 538394 | 167588 |
| 5 | 254172 | 184867 | 26544 | 126533 | 251503 | 252217 | 364119 | 89003 | 72914 | 362469 |
| 6 | 142978 | 173349 | 138970 | 20175 | 98063 | 165219 | 208021 | 244570 | 87323 | 48696 |
| 7 | 145385 | 116328 | 112476 | 90151 | 22086 | 62363 | 126174 | 150588 | 201021 | 58116 |
| 8 | 54778 | 125548 | 89672 | 72031 | 61813 | 19562 | 42569 | 85863 | 122496 | 111251 |
| 9 | 130771 | 41186 | 88726 | 48668 | 47925 | 47560 | 13533 | 34795 | 55913 | 68240 |
| 10 | 39920 | 146186 | 27552 | 49252 | 37482 | 37607 | 32786 | 19658 | 20710 | 32228 |
| 11 | 56210 | 31639 | 91743 | 19745 | 30105 | 26965 | 22971 | 25747 | 13178 | 13904 |
| 12 | 104927 | 199615 | 156121 | 132040 | 69183 | 97652 | 81153 | 63146 | 57494 | 35814 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 24246 | 10007 | 43447 | 19354 | 25368 | 14759 | 37956 | 36012 | 61127 | 67003 |
| 1 | 140534 | 58459 | 83583 | 128144 | 147315 | 81529 | 119852 | 144390 | 99352 | 73597 |
| 2 | 209848 | 212521 | 156292 | 210319 | 221489 | 340898 | 168882 | 186481 | 229767 | 132994 |
| 3 | 410751 | 206421 | 356209 | 266677 | 306979 | 340215 | 333365 | 238426 | 264566 | 223639 |
| 4 | 208146 | 375451 | 266591 | 398240 | 267420 | 275031 | 279182 | 378881 | 323186 | 261778 |
| 5 | 156742 | 188623 | 306143 | 244285 | 301346 | 186855 | 177667 | 246781 | 361945 | 281041 |
| 6 | 254015 | 129145 | 156070 | 255472 | 184925 | 197856 | 96303 | 135059 | 207619 | 244212 |
| 7 | 42549 | 197888 | 113899 | 149932 | 189847 | 142342 | 119831 | 84378 | 118388 | 159019 |
| 8 | 49698 | 51077 | 138458 | 97746 | 106108 | 113413 | 55812 | 66504 | 72745 | 86739 |
| 9 | 85447 | 43415 | 51208 | 121400 | 80054 | 69191 | 59801 | 39450 | 47353 | 50613 |
| 10 | 33041 | 70839 | 36612 | 38794 | 57622 | 42441 | 25803 | 26735 | 24386 | 30363 |
| 11 | 16587 | 29743 | 40956 | 29067 | 20407 | 37960 | 18353 | 13950 | 16551 | 17048 |
| 12 | 27905 | 52986 | 68205 | 68217 | 57551 | 39753 | 30648 | 24974 | 22932 | 32446 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 0 | 36345 | 26034 | 70409 | 14744 | 11553 | 12426 | 75651 | 19302 | 25886 | 17615 |
| 1 | 102407 | 40315 | 222577 | 187997 | 31421 | 46840 | 149425 | 88439 | 59899 | 36514 |
| 2 | 142898 | 158943 | 70041 | 275661 | 453133 | 135648 | 173646 | 190857 | 167748 | 113574 |
| 3 | 275376 | 234186 | 367902 | 91075 | 529753 | 668588 | 159455 | 220575 | 399086 | 455113 |
| 4 | 390858 | 297206 | 350163 | 295777 | 147973 | 293579 | 470063 | 215655 | 284660 | 616963 |
| 5 | 295516 | 309937 | 262716 | 235052 | 258177 | 120538 | 195594 | 455131 | 260314 | 319465 |
| 6 | 241550 | 231804 | 237066 | 183036 | 145899 | 121477 | 97061 | 203492 | 255675 | 224848 |
| 7 | 175608 | 195250 | 151320 | 133595 | 89856 | 63612 | 73510 | 77859 | 124382 | 194326 |
| 8 | 106291 | 120241 | 118870 | 94168 | 65669 | 38763 | 33399 | 59652 | 57297 | 73171 |
| 9 | 52394 | 72205 | 79945 | 75701 | 40443 | 23947 | 18961 | 30494 | 32343 | 29738 |
| 10 | 31280 | 42529 | 43789 | 45951 | 35654 | 18612 | 13987 | 16039 | 19482 | 14989 |
| 11 | 18918 | 20546 | 21611 | 25797 | 16430 | 7955 | 8334 | 11416 | 6798 | 7470 |
| 12 | 34202 | 40706 | 40280 | 30890 | 19509 | 10669 | 10186 | 12801 | 9581 | 5003 |


| age | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| ---: | :--- | ---: | :--- | ---: | :--- | ---: | :--- | ---: | ---: | ---: | ---: | ---: |
| 0 | 23453 | 30429 | 23877 | 11325 | 62142 | 6732 | 716 | 28306 | 6995 | 6236 |
| 1 | 78636 | 62748 | 66370 | 47077 | 44558 | 104282 | 57466 | 43763 | 40332 | 41921 |
| 2 | 137351 | 115701 | 204121 | 235494 | 138880 | 127940 | 205840 | 89101 | 236207 | 126073 |
| 3 | 304647 | 323847 | 216711 | 400036 | 672022 | 250575 | 258176 | 461621 | 136779 | 350611 |
| 4 | 740816 | 471564 | 417953 | 371713 | 832975 | 583694 | 427212 | 353230 | 376312 | 114606 |
| 5 | 613418 | 656507 | 458718 | 445515 | 568835 | 651786 | 593046 | 398273 | 257069 | 295731 |
| 6 | 285438 | 490219 | 514489 | 433533 | 554367 | 453084 | 534943 | 505073 | 294539 | 226640 |
| 7 | 143537 | 244725 | 325982 | 340686 | 506804 | 416897 | 341408 | 432242 | 424715 | 229725 |
| 8 | 102446 | 113277 | 143643 | 190660 | 341618 | 356936 | 270586 | 262799 | 316779 | 267491 |
| 9 | 45963 | 53512 | 69962 | 113220 | 142398 | 206045 | 170574 | 189449 | 197761 | 204818 |
| 10 | 21268 | 25081 | 30761 | 46269 | 63871 | 107830 | 94849 | 138347 | 140403 | 102991 |
| 11 | 6272 | 12322 | 11657 | 19025 | 21501 | 26978 | 33910 | 59278 | 82812 | 66976 |
| 12 | 8529 | 10792 | 11720 | 17890 | 14123 | 22741 | 24427 | 51139 | 60485 | 74918 |

```
    year
    2020
        6443
        52637
    107302
    182163
    266760
    166627
    270154
    246268
    274182
    311215
    241775
    128294
    1 7 9 7 0 3
```


## Table 8.7.1.3. NE Atlantic Mackerel. WEIGHTS AT AGE IN THE CATCH

```
Units : Kg
    year
age 19800
    0}00.057 0.060 0.053 0.050 0.031 0.055 0.039 0.076 0.055 0.049 0.085 0.068
    1}0.131 0.132 0.131 0.168 0.102 0.144 0.146 0.179 0.133 0.136 0.156 0.156
```



```
    0.285 0.287 0.285 0.276 0.295 0.357 0.335 0.318 0.323 0.320 0.336 0.327
    0.345 0.344 0.345 0.310 0.326 0.418 0.423 0.399 0.388 0.377 0.379 0.394
    0.378 0.377 0.378 0.386 0.344 0.417 0.471 0.474 0.456 0.433 0.423 0.423
    llllllllllllllllll
    llllllllllllllllllll
```



```
    0.542 0.543 0.541 0.545 0.569 0.564 0.591 0.580 0.562 0.578 0.606 0.609
    0.574 0.573 0.574 0.606 0.628 0.629 0.552 0.634 0.613 0.581 0.606 0.630
    0.590 0.576 0.574 0.608 0.636 0.679 0.694 0.635 0.624 0.648 0.591 0.649
    0.580 0.584 0.582 0.614 0.663 0.710 0.688 0.718 0.697 0.739 0.713 0.708
        year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003
    0.051 0.061 0.046 0.072 0.058 0.076 0.065 0.062 0.063 0.069 0.052 0.081
    0.167 0.134 0.136 0.143 0.143 0.143 0.157 0.176 0.135 0.172 0.160 0.170
```

```
    2 0.239 0.240 0.255 0.234 0.226 0.230 0.227 0.235 0.227 0.224 0.256 0.267
    3 0.333 0.317 0.339 0.333 0.313 0. 295 0.310 0.306 0.306 0.305 0.307 0.336
    0.397 0.376 0.390 0.390 0.377 0.359 0.354 0.361 0.363 0.376 0.368 0.385
    llllllllllllllllll}0.460.436 0.448 0.452 0.425 0.415 0.408 0.404 0.427 0.424 0.424 0.438
    llllllllllllllllllll
    0.532 0.527 0.543 0.539 0.518 0.481 0.462 0.500 0.501 0.496 0.512 0.522
    0.555 0.548 0.590 0.577 0.551 0.524 0.518 0.536 0.534 0.540}0.5.536 0.572
```



```
    0.651 0.595 0.627 0.606 0.596 0.577 0.573 0.586 0.586 0.603 0.600 0.631
    0.663 0.647 0.678 0.631 0.603 0.591 0.591 0.607 0.594 0.611 0.629 0.648
    2 0.669 0.679 0.713 0.672 0.670 0.636 0.631 0.687 0.644 0.666 0.665 0.715
    year
age 2004 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015
    0.067 0.048 0.038 0.089 0.051 0.104 0.048 0.029 0.089 0.091 0.043 0.051
    0.156 0.151 0.071 0.120 0.105 0.153 0.118 0.113 0.123 0.173 0.126 0.154
    0.263 0.268 0.197 0.215 0.222 0.213 0.221 0.231 0.186 0.234 0.231 0.242
    0.323 0.306 0.307 0.292 0.292 0.283 0.291 0.282 0.284 0.277 0.282 0.294
    0.400 0.366 0.357 0.372 0.370 0.331 0.331 0.334 0.340 0.336 0.324 0.320
    0.419 0.434 0.428 0.408 0.418 0.389 0.365 0.368 0.374 0.360}00.3620.351
    0.485 0.440 0.479 0.456 0.444 0.424 0.418 0.411 0.401 0.386 0.394 0.392
    lllllllllllllllllllll
    0.554 0.539 0.543 0.534 0.551 0.497 0.487 0.494 0.469 0.431 0.443 0.443
    0.573 0.556 0.584 0.573 0.571 0.538 0.515 0.540}0.50.503 0.454 0.467 0.465
    0.595 0.583 0.625 0.571 0.620 0.586 0.573 0.580 0.537 0.472 0.482 0.489
    0.630 0.632 0.636 0.585 0.595 0.599 0.603 0.611 0.537 0.493 0.523 0.522
    0.684 0.655 0.689 0.666 0.662 0.630}0.6.630 0.664 0.585 0.554 0.589 0.561
```


## year

age 20162017201820192020
$0.0350 .0180 .0660 .057 \quad 0.057$
$\begin{array}{lllll}0.154 & 0.178 & 0.147 & 0.112 & 0.174\end{array}$
$0.240 \quad 0.2660 .247 \quad 0.2600 .285$
$0.2970 .3110 .320 \quad 0.2970 .322$
$0.3290 .3560 .3550 .360 \quad 0.360$
$0.3560 .377 \quad 0.397 \quad 0.388 \quad 0.389$
$\begin{array}{lllll}0.383 & 0.397 & 0.410 & 0.429 & 0.417\end{array}$
$0.4110 .415 \quad 0.426 \quad 0.441 \quad 0.444$
$\begin{array}{lllll}0.438 & 0.444 & 0.446 & 0.453 & 0.459\end{array}$
$\begin{array}{lllll} & 0.453 & 0.465 & 0.469 & 0.472\end{array} 0.471$
$\begin{array}{llllll}10 & 0.479 & 0.484 & 0.492 & 0.497 & 0.495\end{array}$
$\begin{array}{llllll}1 & 0.499 & 0.497 & 0.507 & 0.514 & 0.519\end{array}$
$\begin{array}{llllll}2 & 0.520 & 0.531 & 0.537 & 0.537 & 0.554\end{array}$

## Table 8.7.1.4. NE Atlantic Mackerel. WEIGHTS AT AGE IN THE STOCK

```
Units : Kg
    year
age 1980
    0.063 0.063 0.063 0.063 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.114 0.112 0.112 0.111 0.108 0.111 0.104 0.075 0.099 0.058 0.096 0.174
    0.205 0.179 0.159 0.179 0.204 0.244 0.184 0.157 0.181 0.162 0.166 0.184
    0.287 0.258 0.217 0.233 0.251 0.281 0.269 0.234 0.238 0.230 0.247 0.243
    0.322 0.312 0.300 0.282 0.293 0.308 0.301 0.318 0.298 0.272 0.290 0.303
    0.356 0.335 0.368 0.341 0.326 0.336 0.350 0.368 0.348 0.338 0.332 0.347
    0.377 0.376 0.362 0.416 0.395 0.356 0.350 0.414 0.392 0.392 0.383 0.392
    0.402 0.415 0.411 0.404 0.430 0.407 0.374 0.415 0.445 0.388 0.435 0.423
    0.434 0.431 0.456 0.438 0.455 0.455 0.434 0.431 0.442 0.449 0.447 0.492
    9 0.438 0.454 0.455 0.475 0.489 0.447 0.428 0.483 0.466 0.432 0.494 0.500
    10 0.484 0.450 0.473 0.467 0.507 0.519 0.467 0.487 0.506 0.429 0.473 0.546
    11 0.520 0.524 0.536 0.544 0.513 0.538 0.506 0.492 0.567 0.482 0.495 0.526
    12 0.532 0.530 0.542 0.528 0.566 0.590 0.541 0.581 0.594 0.556 0.536 0.619
        year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003
    0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.130 0.145 0.114 0.116 0.097 0.084 0.083 0.087 0.093 0.113 0.109 0.112
    0.201 0.190 0.163 0.200 0.185 0.196 0.170 0.210 0.194 0.190 0.206 0.181
    0.260 0.266 0.240 0.278 0.250 0.257 0.251 0.260 0.253 0.246 0.245 0.251
    0.308 0.323 0.306 0.327 0.322 0.310 0.300 0.317 0.301 0.303 0. 0.388 0.277
    0.360 0.359 0.368 0.385 0.372 0.356 0.348 0.356 0.357 0.342 0.333 0.341
    0.397 0.410 0.418 0.432 0.425 0.401 0.384 0.392 0.394 0.398 0.360 0.401
    0.419 0.432 0.459 0.458 0.446 0.460 0.409 0.424 0.415 0.417 0.418 0.407
    llllllllllllllllllll
    0.487 0.480 0.496 0.511 0.513 0.505 0.475 0.489 0.464 0.484 0.458 0.490
    lllllllllllllllllllll
    10.543 0.547 0.592 0.560 0.538}0.5.546 0.500 0.545 0.514 0.535 0.523 0.521
    2 0.572 0.580 0.608 0.603 0.573 0.583 0.549 0.575 0.551 0.572 0.558 0.540
        year
age 2004 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015
    0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.112 0.114 0.114 0.095 0.133 0.112 0.096 0.080 0.089 0.076 0.107 0.078
    0.157 0.140 0.164 0.148 0.160 0.162 0.159 0.175 0.155 0.144 0.165 0.207
    0.258 0.221 0.236 0.206 0.207 0.214 0.199 0.223 0.216 0.179 0.199 0.247
    0.319 0.328 0.291 0.285 0.260 0.268 0.246 0.274 0.255 0.249 0.238}00.254
    0.356 0.378 0.333 0.329 0.346 0.295 0.296 0.332 0.288 0.280}0.30.291 0.288
    0.406 0.403 0.400 0.363 0.354 0.351 0.345 0.369 0.312 0.319 0.321 0.336
    0.449 0.464 0.413 0.448 0.393 0.386 0.389 0.389 0.360 0.341 0.341 0.350
    0.482 0.481 0.437 0.452 0.448 0.437 0.407 0.430}0.40.390 0.375 0.387 0.381
    0.506 0.547 0.455 0.514 0.452 0.461 0.439 0.452 0.453 0.416 0.416 0.412
    0.519 0.538 0.469 0.538}0.4.478 0.517 0.489 0.495 0.498 0.441 0.466 0.447
    0.579 0.509 0.531 0.542 0.487 0.548 0.532 0.518 0.503 0.496 0.472 0.485
    0.588}0.6030.566 0.585 0.510 0.557 0.572 0.525 0.558 0.522 0.517 0.551
        year
age 2016 2017 2018 2019 2020
    0.000 0.000 0.000 0.000 0.000
    0.059 0.058 0.064 0.070 0.068
    0.182 0.204 0.190 0.191 0.210
    0.238 0.237 0.266 0.250 0.252
```

```
4 0.282 0.278 0.283 0.293 0.289
5}00.298 0.308 0.314 0.311 0.348
6 0.340 0.308 0.327 0.346 0.363
7 0.368 0.338 0.346 0.365 0.375
8 0.385 0.377 0.364 0.371 0.394
9 0.404 0.394 0.389 0.397 0.400
10}00.424 0.426 0.419 0.428 0.423
11}00.440 0.430 0.437 0.431 0.445
12 0.473 0.499 0.491 0.481 0.486
```


## Table 8.7.1.5. NE Atlantic Mackerel. NATURAL MORTALITY

```
Units : NA
    year
age 1980 1981 1982 1983 1984 1985 1986 1987 1988 1989 1990 1991 1992 1993 1994
    0}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    1 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    2 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    3 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    4 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    5 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    6 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    7 0.15}0.1
```



```
    9 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```





```
        year
age 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007 2008 2009
```







```
    5
    6
```



```
    8
    9}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```




```
    12}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
        year
age 2010 2011 2012 2013 2014 2015 2016 2017 2018 20192020
    0.15}0.150.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    0.15}0.150.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    4 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    5 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    6 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    7}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```



```
9 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
10}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
11}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```



## Table 8.7.1.6. NE Atlantic Mackerel. PROPORTION MATURE

$\begin{array}{llllllllllllll}\text { age } & 1980 & 1981 & 1982 & 1983 & 1984 & 1985 & 1986 & 1987 & 1988 & 1989 & 1990 & 1991\end{array}$
$00.0000 .0000 .0000 .0000 .0000 .000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000$ $\begin{array}{lllllllllllllllll}0.093 & 0.097 & 0.097 & 0.098 & 0.102 & 0.102 & 0.102 & 0.102 & 0.102 & 0.102 & 0.102 & 0.102\end{array}$ $\begin{array}{lllllllllllll}0.521 & 0.497 & 0.498 & 0.485 & 0.467 & 0.516 & 0.522 & 0.352 & 0.360 & 0.372 & 0.392 & 0.435\end{array}$ $\begin{array}{llllllllllllll}0.872 & 0.837 & 0.857 & 0.863 & 0.853 & 0.885 & 0.926 & 0.922 & 0.901 & 0.915 & 0.909 & 0.912\end{array}$ $0.9490 .9340 .9300 .9400 .9380 .940 \quad 0.9830 .9940 .9890 .9940 .9960 .991$ $\begin{array}{llllllllllllllll}0.972 & 0.976 & 0.969 & 0.972 & 0.966 & 0.966 & 0.965 & 0.997 & 0.994 & 0.996 & 0.998 & 0.996\end{array}$ 0.9840 .9840 .9870 .9991 .0001 .0001 .0001 .0001 .0001 .0001 .00010 .996 $0.990 \quad 0.9870 .9850 .9840 .9750 .9761 .0001 .0001 .0001 .0001 .0001 .000$ $\begin{array}{llllllllllllllllll}1.000 & 0.999 & 0.999 & 0.999 & 0.999 & 0.999 & 0.991 & 0.992 & 0.991 & 0.993 & 0.995 & 1.000\end{array}$ 1.0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .000 $\begin{array}{llllllllllllll}1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.0001 .0001 .0001 .000\end{array}$ 11.0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .000 $\begin{array}{llllllllllllll}2 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.0001 .0001 .000\end{array}$ year
age 199219931994199519961997199819992000200120022003
$0.0000 .0000 .0000 .0000 .0000 .0000 .000 \quad 0.0000 .0000 .000 \quad 0.0000 .000$
$\begin{array}{llllllllllllllllll}0.102 & 0.102 & 0.102 & 0.102 & 0.102 & 0.097 & 0.097 & 0.097 & 0.104 & 0.104 & 0.104 & 0.106\end{array}$
$\begin{array}{llllllllllllll}0.520 & 0.534 & 0.621 & 0.599 & 0.586 & 0.621 & 0.688 & 0.669 & 0.692 & 0.675 & 0.710 & 0.690\end{array}$
$\begin{array}{lllllllllllllll}0.928 & 0.934 & 0.938 & 0.931 & 0.936 & 0.880 & 0.886 & 0.876 & 0.909 & 0.909 & 0.937 & 0.940\end{array}$
$\begin{array}{lllllllllllllllll}0.996 & 0.996 & 0.994 & 0.993 & 1.000 & 0.993 & 0.994 & 0.989 & 0.989 & 0.987 & 0.992 & 0.988\end{array}$
$\begin{array}{lllllllllllllll}0.997 & 0.997 & 0.997 & 0.994 & 1.000 & 0.998 & 0.999 & 0.999 & 0.998 & 0.998 & 1.000 & 1.000\end{array}$
$\begin{array}{llllllllllllllll}0.994 & 0.994 & 0.993 & 0.987 & 0.994 & 0.999 & 0.999 & 0.999 & 0.999 & 0.999 & 1.000 & 1.000\end{array}$
$\begin{array}{lllllllllllllll}1.000 & 1.000 & 0.999 & 0.999 & 0.999 & 1.000 & 1.000 & 1.000 & 1.000 & 0.999 & 1.000 & 0.999\end{array}$
$\begin{array}{lllllllllllll}1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 0.994 & 0.995 & 0.996 & 0.997 & 0.997 & 1.000 & 1.000\end{array}$
$\begin{array}{lllllllllllll}1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000\end{array}$
$\begin{array}{llllllllllllllll}1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.0001 .000\end{array}$
$\begin{array}{llllllllllllll}1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000\end{array}$
$\begin{array}{llllllllllllllllll}2 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.0001 .0001 .0001 .000\end{array}$
year
age $2004 \quad 2005 \quad 2006 \quad 2007 \quad 2008 \quad 2009 \quad 2010 \quad 2011 \quad 2012 \quad 2013 \quad 2014 \quad 2015$
$\begin{array}{lllllllllllll}0.000 & 0.000 & 0.000 & 0.000 & 0.000 & 0.000 & 0.000 & 0.000 & 0.000 & 0.000 \quad 0.000 \quad 0.000\end{array}$
$\begin{array}{llllllllllllll}0.106 & 0.106 & 0.095 & 0.095 & 0.095 & 0.096 & 0.096 & 0.096 & 0.094 & 0.092 & 0.092 & 0.104\end{array}$
$\begin{array}{lllllllllllllllll}0.761 & 0.616 & 0.589 & 0.546 & 0.524 & 0.541 & 0.667 & 0.655 & 0.604 & 0.683 & 0.675 & 0.763\end{array}$
$\begin{array}{lllllllllllll}0.962 & 0.959 & 0.928 & 0.921 & 0.917 & 0.919 & 0.930 & 0.927 & 0.926 & 0.921 & 0.916 & 0.944\end{array}$
$\begin{array}{lllllllllllll}0.993 & 0.993 & 0.994 & 0.994 & 0.999 & 0.999 & 0.999 & 0.999 & 0.999 & 0.998 & 0.999 & 0.998\end{array}$
$0.9990 .9991 .0001 .0000 .9991 .0001 .0001 .000 \quad 0.9991 .0001 .000 \quad 0.999$
$\begin{array}{llllllllllllll}1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 0.999 & 0.999 & 0.999 & 0.999 & 0.999 & 1.000\end{array}$
$0.9990 .9991 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .000 \quad 0.999 \quad 0.999$
1.0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .000
$1.0001 .0001 .0001 .0001 .000 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.0001 .0001 .000$
$1.0001 .0001 .0001 .0001 .0001 .0001 .000 \quad 1.000 \quad 1.0001 .0001 .0001 .000$
1.0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .000
1.0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .000

```
    year
age 2016 2017 2018 2019 2020
0 0.000 0.000 0.000 0.000 0.000
10.111 0.109 0.092 0.092 0.092
2 0.632 0.604 0.469 0.440 0.420
30.937 0.945 0.902 0.902 0.909
0.997 0.998 0.999 0.998 0.998
0.999 1.000 1.000 1.000 1.000
1.000 1.000 1.000 1.000 0.999
0.999 0.999 0.999 1.000 0.999
1.000 1.000 1.000 1.000 1.000
1.000 1.000 1.000 1.000 1.000
10 1.000 1.000 1.000 1.000 1.000
11 1.000 1.000 1.000 1.000 1.000
12 1.000 1.000 1.000 1.000 1.000
```


## Table 8.7.1.7. NE Atlantic Mackerel. FRACTION OF HARVEST BEFORE SPAWNING

| year |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.139 | 0.111 |
| 2 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.240 | 0.272 |
| 3 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.240 | 0.272 |
| 4 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 | 0.240 | 0.272 |
| 5 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.393 | 0.406 |
| 6 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.393 | 0.406 |
| 7 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.393 | 0.406 |
| 8 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.393 | 0.406 |
| 9 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.393 | 0.406 |
| 10 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.393 | 0.406 |
| 11 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.393 | 0.406 |
| 12 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.393 | 0.406 |

    year
    age $1992199319941995 \quad 19961997 \quad 1998 \quad 1999 \quad 2000 \quad 2001 \quad 2002 \quad 2003$
$0.0000 .0000 .0000 .0000 .000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000$
$\begin{array}{llllllllllllll}0.084 & 0.165 & 0.249 & 0.331 & 0.269 & 0.206 & 0.144 & 0.125 & 0.106 & 0.088 & 0.142 & 0.197\end{array}$
$\begin{array}{lllllllllllllll}0.304 & 0.301 & 0.298 & 0.296 & 0.295 & 0.295 & 0.295 & 0.320 & 0.347 & 0.373 & 0.360 & 0.347\end{array}$
$\begin{array}{lllllllllllllllll}0.304 & 0.301 & 0.298 & 0.296 & 0.295 & 0.295 & 0.295 & 0.320 & 0.347 & 0.373 & 0.360 & 0.347\end{array}$
$\begin{array}{lllllllllllllllll}0.304 & 0.301 & 0.298 & 0.296 & 0.295 & 0.295 & 0.295 & 0.320 & 0.347 & 0.373 & 0.360 & 0.347\end{array}$
$\begin{array}{lllllllllllll}0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$\begin{array}{lllllllllllll}0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$\begin{array}{llllllllllll}0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$\begin{array}{llllllllllll}0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$\begin{array}{llllllllllll}0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$\begin{array}{llllllllllll}0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$\begin{array}{lllllllllllll}0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$\begin{array}{llllllllllllll}0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
year
age 2004 2005 2006 2007 $2008 \quad 2009 \quad 2010 \quad 2011 \quad 2012 \quad 2013 \quad 2014 \quad 2015$
$0.0000 .0000 .0000 .0000 .0000 .0000 .0000 .000 \quad 0.000 \quad 0.000 \quad 0.0000 .000$
$\begin{array}{lllllllllllllllll}0.251 & 0.262 & 0.274 & 0.285 & 0.206 & 0.125 & 0.047 & 0.092 & 0.138 & 0.183 & 0.170 & 0.156\end{array}$
$\begin{array}{lllllllllllllll}0.334 & 0.317 & 0.300 & 0.284 & 0.266 & 0.249 & 0.232 & 0.176 & 0.119 & 0.064 & 0.117 & 0.171\end{array}$
$\begin{array}{llllllllllllllllll}0.334 & 0.317 & 0.300 & 0.284 & 0.266 & 0.249 & 0.232 & 0.176 & 0.119 & 0.064 & 0.117 & 0.171\end{array}$

```
4 0.334 0.317 0.300 0.284 0.266 0.249 0.232 0.176 0.119 0.064 0.117 0.171
5 0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
0}0.4410.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
12 0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
    year
age 2016 2017 2018 2019 2020
    0.000 0.000 0.000 0.000 0.000
    0.143 0.232 0.393 0.581 0.532
    0.224 0.153 0.180 0.183 0.184
    0.224 0.153 0.180 0.183 0.184
    0.224 0.153 0.180 0.183 0.184
    0.176 0.291 0.193 0.299 0.321
    0.176 0.291 0.193 0.299 0.321
    0.176 0.291 0.193 0.299 0.321
    0.176 0.291 0.193 0.299 0.321
    0.176 0.291 0.193 0.299 0.321
    0.176 0.291 0.193 0.299 0.321
    0.176 0.291 0.193 0.299 0.321
    20.176 0.291 0.193 0.299 0.321
```

Table 8.7.1.8. NE Atlantic Mackerel. FRACTION OF NATURAL MORTALITY BEFORE SPAWNING
year
$\begin{array}{llllllllllllll}\text { age } & 1980 & 1981 & 1982 & 1983 & 1984 & 1985 & 1986 & 1987 & 1988 & 1989 & 1990 & 1991\end{array}$
$\begin{array}{lllllllllllllll}0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{lllllllllllllll}0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{llllllllllllllllll}0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{llllllllllllllllll}0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{llllllllllllllllll}0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{lllllllllllllllll}0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{lllllllllllllllll}0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{lllllllllllllll}0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{lllllllllllllll}0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{lllllllllllllllll}0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{llllllllllllllllllll}0 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{lllllllllllllllllll}1 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{lllllllllllllllllllll}12 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
year
age
$\begin{array}{llllllllllllll}0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{lllllllllllllllllllll}0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{lllllllllllllllll}0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{llllllllllllll}0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{llllllllllllllllll}0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{lllllllllllllllllll}0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{llllllllllllllllllll}0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{lllllllllllllll}0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{lllllllllllllllllll}0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{lllllllllllllllllll}9 & 0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$

```
    10 0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    11 0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325}0.3.346 0.366 0.361 0.355
    12 0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
        year
age 2004 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015
    0}0.3500.346 0.342 0.339 0.311 0.383 0.255 0.252 0.249 0.246 0.278 0.311
    1}00.3500.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
    2 0.350 0.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
    3 0.350 0.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
    4 0.350}00.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
    5 0.350 0.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
    6
    7 0.350 0.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
    8}00.350 0.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
    9}00.3500.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
    10}00.350 0.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
    11 0.350 0.346 0.342 0.339}0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
    12 0.350 0.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
        year
age 2016 2017 2018 2019 2020
    00.343 0.327 0.312 0.296 0.296
    10.343 0.327 0.312 0.296 0.296
    20.343 0.327 0.312 0.296 0.296
    30.343 0.327 0.312 0.296 0.296
    40.343 0.327 0.312 0.296 0.296
    5 0.343 0.327 0.312 0.296 0.296
    6 0.343 0.327 0.312 0.296 0.296
    70.343 0.327 0.312 0.296 0.296
    80.343 0.327 0.312 0.296 0.296
    90.343 0.327 0.312 0.296 0.296
    10 0.343 0.327 0.312 0.296 0.296
    11 0.343 0.327 0.312 0.296 0.296
    120.343 0.327 0.312 0.296 0.296
```


## Table 8.7.1.9. NE Atlantic Mackerel. SURVEY INDICES

Some random text

SSB-egg-based-survey
19922020
$-1 \quad-1$
3782966.707
-1
-1
4810751.571
-1
-1
4831948.353
-1
-1
3524054.85
1 -1

| 1 | -1 |  |  |
| :---: | :---: | :---: | :---: |
| 1 | 3087517.078 |  |  |
| 1 | -1 |  |  |
| R-idx |  |  |  |
| 1998 | 2020 |  |  |
| 1 | 1 | 0 | 0 |
| 0 | 0 |  |  |
| 1 | 0.012476066 |  |  |
| 1 | 0.01862673 |  |  |
| 1 | 0.013289745 |  |  |
| 1 | 0.020583855 |  |  |
| 1 | 0.026244937 |  |  |
| 1 | 0.012684229 |  |  |
| 1 | 0.029582367 |  |  |
| 1 | 0.038157763 |  |  |
| 1 | 0.034722557 |  |  |
| 1 | 0.022670008 |  |  |
| 1 | 0.02064922 |  |  |
| 1 | 0.014607073 |  |  |
| 1 | 0.02237237 |  |  |
| 1 | 0.037563703 |  |  |
| 1 | 0.02733911 |  |  |
| 1 | 0.029964112 |  |  |
| 1 | 0.022348323 |  |  |
| 1 | 0.024720467 |  |  |
| 1 | 0.0432534 |  |  |
| 1 | 0.043849281 |  |  |
| 1 | 0.039094593 |  |  |
| 1 | 0.04381569 |  |  |
| 1 | 0.036397234 |  |  |
| Swept-idx |  |  |  |
| 2010 | 2021 |  |  |
| 1 | 1 | 0.58 | 0.75 |
| 3 | 11 |  |  |

Table 8.7.1.10. NE Atlantic Mackerel. RFID recapture data for the year 2020.

| Release Yr | Recapture Yr | Year- <br> class | age at release | Numbers scanned in recapture Yr | Numbers Released in Release Year | Numbers recaptured |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2018 | 2020 | 2007 | 19391477 | 1670.4499 | 7 7 |  |
| 2018 | 2020 | 2008 | 29244736 | 4092.9627 | 202 |  |
| 2018 | 2020 | 2009 | 39505301 | 3273.9251 | 17 2 |  |
| 2018 | 2020 | 2010 | 99081840 | 6506.48 | 40 2 |  |
| 2018 | 2020 | 2011 | 110470858 | 7923.5647 | 50 2 |  |
| 2018 | 2020 | 2012 | 61620787 | 2290.2767 | 15 2 |  |
| 2018 | 2020 | 2013 | 53083627 | 3049.499 | 202 |  |
| 2019 | 2020 | 2008 | 29244736 | 2556.359 | 28 2 |  |
| 2019 | 2020 | 2009 | 39505301 | 2871.3265 | 30 2 |  |
| 2019 | 2020 | 2010 | 99081840 | 4727.5524 | 49 2 |  |
| 2019 | 2020 | 2011 | 110470858 | 9482.5831 | 101 2 |  |
| 2019 | 2020 | 2012 | 61620787 | 6784.5181 | 72 2 |  |
| 2019 | 2020 | 2013 | 53083627 | 8039.9448 | 82 2 |  |
| 2019 | 2020 | 2014 | 73636345 | 5824.132 | 592 |  |

Table 8.7.2.1. NE Atlantic Mackerel. SAM parameter estimates for the 2021 update.

|  | esti- <br> mate | std.dev | confidence interval lower <br> bound | confidence interval upper <br> bound |
| :--- | :--- | :--- | :--- | :--- |
| observation standard deviations |  |  |  |  |
| Catches age 0 | 0.91 | 0.18 | 0.63 | 1.29 |
| Catches age 1 | 0.36 | 0.23 | 0.23 | 0.58 |
| Catches age 2-12 | 0.11 | 0.16 | 0.08 | 0.15 |
| Egg survey | 0.31 | 0.26 | 0.19 | 0.50 |
| Recruitment index | 0.28 | 0.30 | 0.15 | 1.05 |
| IESSNS age 3 | 0.65 | 0.24 | 0.40 | 1.14 |
| IESSNS ages 4-11 | 0.39 | 0.14 | 0.29 | 0.51 |
| Recapture overdispersion <br> tags | 1.23 | 0.25 | 1.38 | 0.53 |


|  | estimate | std.dev | confidence interval lower bound | confidence interval upper bound |
| :---: | :---: | :---: | :---: | :---: |
| random walk standard deviation |  |  |  |  |
| F age 0 | 0.25 | 0.49 | 0.09 | 0.66 |
| F age 1 | 0.15 | 0.49 | 0.06 | 0.40 |
| F age 2+ | 0.13 | 0.19 | 0.09 | 0.18 |
| N@age0 | 0.16 | 0.74 | 0.04 | 0.70 |
| process error standard deviation |  |  |  |  |
| N@age1-12+ | 0.21 | 0.09 | 0.18 | 0.26 |
| catchabilities |  |  |  |  |
| egg survey | 1.22 | 0.11 | 0.98 | 1.53 |
| recruitment index | $\begin{aligned} & 5.13 \mathrm{E}- \\ & 09 \end{aligned}$ | $\begin{aligned} & 1.25 \mathrm{E}- \\ & 01 \end{aligned}$ | 3.99E-09 | 6.59E-09 |
| IESSNS age 3 | 0.82 | 0.23 | 0.52 | 1.30 |
| IESSNS age 4 | 1.25 | 0.16 | 0.91 | 1.74 |
| IESSNS age 5 | 1.71 | 0.16 | 1.24 | 2.37 |
| IESSNS age 6 | 1.83 | 0.16 | 1.32 | 2.53 |
| IESSNS age 7 | 1.95 | 0.16 | 1.41 | 2.70 |
| IESSNS age 8 | 1.85 | 0.16 | 1.34 | 2.56 |
| IESSNS age 9 | 1.95 | 0.16 | 1.41 | 2.69 |
| IESSNS ages 10-11 | 1.76 | 0.16 | 1.28 | 2.42 |
| post tagging survival steal tags | 0.40 | 0.11 | 0.35 | 0.46 |
| post tagging survival RFID tags | 0.15 | 0.11 | 0.12 | 0.18 |

Table 8.7.3.1. NE Atlantic Mackerel. STOCK SUMMARY.

| Year |  | Recruitment | SSB |  | Total |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| Year$1997$ |  | Recruitment |  | SSB |  | Total |  |  | $0.36$ | 0.25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & 299315 \\ & 2 \end{aligned}$ | $\begin{aligned} & 435777 \\ & 6 \end{aligned}$ | $\begin{aligned} & 205585 \\ & 7 \end{aligned}$ | $\begin{aligned} & 215819 \\ & 9 \end{aligned}$ | $\begin{aligned} & 254215 \\ & 5 \end{aligned}$ | $183223$ | 573029 | 0.30 |  |  |
| 1998 | $\begin{aligned} & 304876 \\ & 1 \end{aligned}$ | $\begin{aligned} & 431951 \\ & 7 \end{aligned}$ | $\begin{aligned} & 215184 \\ & 8 \end{aligned}$ | $\begin{aligned} & 213081 \\ & 7 \end{aligned}$ | $\begin{aligned} & 251710 \\ & 7 \end{aligned}$ | $\begin{aligned} & 180380 \\ & 9 \end{aligned}$ | 666316 | 0.31 | 0.37 | 0.26 |
| 1999 | $\begin{aligned} & 329397 \\ & 4 \end{aligned}$ | $\begin{aligned} & 445035 \\ & 8 \end{aligned}$ | $\begin{aligned} & 243806 \\ & 5 \end{aligned}$ | $\begin{aligned} & 232039 \\ & 4 \end{aligned}$ | $\begin{aligned} & 273916 \\ & 8 \end{aligned}$ | $\begin{aligned} & 196564 \\ & 3 \end{aligned}$ | 640309 | 0.32 | 0.38 | 0.27 |
| 2000 | $\begin{aligned} & 327144 \\ & 9 \end{aligned}$ | $\begin{aligned} & 495128 \\ & 7 \end{aligned}$ | $\begin{aligned} & 216153 \\ & 5 \end{aligned}$ | $\begin{aligned} & 229503 \\ & 0 \end{aligned}$ | $\begin{aligned} & 265034 \\ & 5 \end{aligned}$ | $\begin{aligned} & 198735 \\ & 0 \end{aligned}$ | 738606 | 0.33 | 0.38 | 0.29 |
| 2001 | $\begin{aligned} & 429277 \\ & 7 \end{aligned}$ | $\begin{aligned} & 612657 \\ & 0 \end{aligned}$ | $\begin{aligned} & 300787 \\ & 2 \end{aligned}$ | $\begin{aligned} & 217926 \\ & 1 \end{aligned}$ | $\begin{aligned} & 251220 \\ & 7 \end{aligned}$ | $\begin{aligned} & 189044 \\ & 0 \end{aligned}$ | 737463 | 0.36 | 0.42 | 0.31 |
| 2002 | $\begin{aligned} & 481542 \\ & 3 \end{aligned}$ | $\begin{aligned} & 741702 \\ & 0 \end{aligned}$ | $\begin{aligned} & 312636 \\ & 3 \end{aligned}$ | $\begin{aligned} & 209099 \\ & 9 \end{aligned}$ | $\begin{aligned} & 243898 \\ & 5 \end{aligned}$ | $\begin{aligned} & 179266 \\ & 2 \end{aligned}$ | 771422 | 0.38 | 0.45 | 0.32 |
| 2003 | $\begin{aligned} & 413090 \\ & 4 \end{aligned}$ | $\begin{aligned} & 609002 \\ & 7 \end{aligned}$ | $\begin{aligned} & 280201 \\ & 8 \end{aligned}$ | $\begin{aligned} & 200863 \\ & 0 \end{aligned}$ | $\begin{aligned} & 234099 \\ & 2 \end{aligned}$ | $\begin{aligned} & 172345 \\ & 5 \end{aligned}$ | 679287 | 0.39 | 0.48 | 0.33 |
| 2004 | $\begin{aligned} & 480278 \\ & 6 \end{aligned}$ | $\begin{aligned} & 658347 \\ & 2 \end{aligned}$ | $\begin{aligned} & 350373 \\ & 6 \end{aligned}$ | $\begin{aligned} & 263309 \\ & 1 \end{aligned}$ | $\begin{aligned} & 311500 \\ & 8 \end{aligned}$ | $\begin{aligned} & 222573 \\ & 0 \end{aligned}$ | 660491 | 0.37 | 0.44 | 0.31 |
| 2005 | $\begin{aligned} & 562564 \\ & 0 \end{aligned}$ | $\begin{aligned} & 918095 \\ & 6 \end{aligned}$ | $\begin{aligned} & 344711 \\ & 6 \end{aligned}$ | $\begin{aligned} & 238211 \\ & 6 \end{aligned}$ | $\begin{aligned} & 282664 \\ & 9 \end{aligned}$ | $\begin{aligned} & 200749 \\ & 2 \end{aligned}$ | 549514 | 0.31 | 0.36 | 0.26 |
| 2006 | $\begin{aligned} & 566755 \\ & 7 \end{aligned}$ | $\begin{aligned} & 898300 \\ & 0 \end{aligned}$ | $\begin{aligned} & 357577 \\ & 6 \end{aligned}$ | $\begin{aligned} & 216683 \\ & 4 \end{aligned}$ | $\begin{aligned} & 256613 \\ & 8 \end{aligned}$ | $\begin{aligned} & 182966 \\ & 4 \end{aligned}$ | 481181 | 0.29 | 0.34 | 0.25 |
| 2007 | $\begin{aligned} & 499574 \\ & 9 \end{aligned}$ | $\begin{aligned} & 684590 \\ & 7 \end{aligned}$ | $\begin{aligned} & 364561 \\ & 0 \end{aligned}$ | $\begin{aligned} & 228890 \\ & 5 \end{aligned}$ | $\begin{aligned} & 269079 \\ & 5 \end{aligned}$ | $194704$ | 586206 | 0.32 | 0.37 | 0.27 |
| 2008 | $\begin{aligned} & 470779 \\ & 2 \end{aligned}$ | $\begin{aligned} & 665103 \\ & 0 \end{aligned}$ | $\begin{aligned} & 333231 \\ & 2 \end{aligned}$ | $\begin{aligned} & 266288 \\ & 3 \end{aligned}$ | $\begin{aligned} & 317433 \\ & 5 \end{aligned}$ | $\begin{aligned} & 223383 \\ & 7 \end{aligned}$ | 623165 | 0.31 | 0.36 | 0.26 |
| 2009 | $\begin{aligned} & 466414 \\ & 1 \end{aligned}$ | $\begin{aligned} & 689240 \\ & 3 \end{aligned}$ | $\begin{aligned} & 315626 \\ & 0 \end{aligned}$ | $\begin{aligned} & 331249 \\ & 4 \end{aligned}$ | $\begin{aligned} & 395939 \\ & 1 \end{aligned}$ | $\begin{aligned} & 277129 \\ & 0 \end{aligned}$ | 737969 | 0.28 | 0.34 | 0.24 |
| 2010 | $\begin{aligned} & 533449 \\ & 9 \end{aligned}$ | $\begin{aligned} & 742473 \\ & 3 \end{aligned}$ | $\begin{aligned} & 383271 \\ & 5 \end{aligned}$ | $\begin{aligned} & 370440 \\ & 9 \end{aligned}$ | $\begin{aligned} & 439847 \\ & 3 \end{aligned}$ | $\begin{aligned} & 311986 \\ & 6 \end{aligned}$ | 877272 | 0.28 | 0.33 | 0.23 |
| 2011 | $\begin{aligned} & 594263 \\ & 3 \end{aligned}$ | $\begin{aligned} & 927328 \\ & 4 \end{aligned}$ | $\begin{aligned} & 380824 \\ & 0 \end{aligned}$ | $\begin{aligned} & 425983 \\ & 5 \end{aligned}$ | $\begin{aligned} & 507608 \\ & 5 \end{aligned}$ | $\begin{aligned} & 357484 \\ & 1 \end{aligned}$ | 948963 | 0.27 | 0.32 | 0.23 |
| 2012 | $\begin{aligned} & 553158 \\ & 2 \end{aligned}$ | $\begin{aligned} & 770552 \\ & 5 \end{aligned}$ | $\begin{aligned} & 397096 \\ & 9 \end{aligned}$ | $\begin{aligned} & 394670 \\ & 7 \end{aligned}$ | $\begin{aligned} & 473882 \\ & 1 \end{aligned}$ | $\begin{aligned} & 328699 \\ & 7 \end{aligned}$ | 899551 | 0.25 | 0.31 | 0.21 |
| 2013 | $\begin{aligned} & 538570 \\ & 7 \end{aligned}$ | $\begin{aligned} & 741974 \\ & 3 \end{aligned}$ | $\begin{aligned} & 390927 \\ & 9 \end{aligned}$ | $\begin{aligned} & 438190 \\ & 9 \end{aligned}$ | $\begin{aligned} & 528930 \\ & 1 \end{aligned}$ | $\begin{aligned} & 363018 \\ & 2 \end{aligned}$ | 938299 | 0.25 | 0.31 | 0.21 |
| 2014 | $\begin{aligned} & 547632 \\ & 9 \end{aligned}$ | $\begin{aligned} & 759792 \\ & 1 \end{aligned}$ | $\begin{aligned} & 394715 \\ & 6 \end{aligned}$ | $\begin{aligned} & 555487 \\ & 0 \end{aligned}$ | $\begin{aligned} & 668612 \\ & 9 \end{aligned}$ | $\begin{aligned} & 461501 \\ & 3 \end{aligned}$ | $\begin{aligned} & 140178 \\ & 8 \end{aligned}$ | 0.26 | 0.31 | 0.21 |
| 2015 | $\begin{aligned} & 517117 \\ & 0 \end{aligned}$ | $\begin{aligned} & 722891 \\ & 7 \end{aligned}$ | $\begin{aligned} & 369917 \\ & 2 \end{aligned}$ | $\begin{aligned} & 555484 \\ & 1 \end{aligned}$ | $673557$ | $\begin{aligned} & 458109 \\ & 0 \end{aligned}$ | $\begin{aligned} & 121582 \\ & 7 \end{aligned}$ | 0.24 | 0.30 | 0.194 |
| 2016 | $\begin{aligned} & 576009 \\ & 4 \end{aligned}$ | $\begin{aligned} & 865592 \\ & 9 \end{aligned}$ | $\begin{aligned} & 383305 \\ & 9 \end{aligned}$ | $\begin{aligned} & 527848 \\ & 1 \end{aligned}$ | $\begin{aligned} & 643292 \\ & 1 \end{aligned}$ | $\begin{aligned} & 433121 \\ & 4 \end{aligned}$ | $\begin{aligned} & 110013 \\ & 5 \end{aligned}$ | 0.22 | 0.27 | 0.174 |


| Year |  | Recruitment |  | SSB |  | Total |  | F |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | $\begin{aligned} & 599028 \\ & 7 \end{aligned}$ | $\begin{aligned} & 930114 \\ & 7 \end{aligned}$ | $\begin{aligned} & 385796 \\ & 9 \end{aligned}$ | $\begin{aligned} & 516163 \\ & 8 \end{aligned}$ | $\begin{aligned} & 629236 \\ & 4 \end{aligned}$ | $\begin{aligned} & 423410 \\ & 2 \end{aligned}$ | $\begin{aligned} & 115964 \\ & 1 \end{aligned}$ | 0.22 | 0.27 | 0.175 |
| 2018 | $\begin{aligned} & 604163 \\ & 6 \end{aligned}$ | $\begin{aligned} & 922964 \\ & 5 \end{aligned}$ | $\begin{aligned} & 395479 \\ & 7 \end{aligned}$ | $\begin{aligned} & 452169 \\ & 1 \end{aligned}$ | $\begin{aligned} & 552948 \\ & 0 \end{aligned}$ | $\begin{aligned} & 369757 \\ & 9 \end{aligned}$ | $\begin{aligned} & 102314 \\ & 4 \end{aligned}$ | 0.22 | 0.27 | 0.175 |
| 2019 | $\begin{aligned} & 651186 \\ & 5 \end{aligned}$ | $\begin{aligned} & 109372 \\ & 79 \end{aligned}$ | $\begin{aligned} & 387705 \\ & 1 \end{aligned}$ | $\begin{aligned} & 387630 \\ & 6 \end{aligned}$ | $\begin{aligned} & 484032 \\ & 8 \end{aligned}$ | $\begin{aligned} & 310428 \\ & 3 \end{aligned}$ | 839727 | 0.22 | 0.27 | 0.170 |
| 2020 | $\begin{aligned} & 574313 \\ & 0^{*} \end{aligned}$ |  |  | $\begin{aligned} & 393855 \\ & 5 \end{aligned}$ | $\begin{aligned} & 501422 \\ & 9 \end{aligned}$ | $\begin{aligned} & 309363 \\ & 9 \end{aligned}$ | $\begin{aligned} & 103951 \\ & 3 \end{aligned}$ | 0.25 | 0.32 | 0.193 |
| 2021 | $\begin{aligned} & 436751 \\ & 3^{* *} \end{aligned}$ |  |  | $\begin{aligned} & 351084 \\ & 9 \dagger \end{aligned}$ |  |  |  |  |  |  |
| Average | $\begin{aligned} & 443722 \\ & 8 \end{aligned}$ | $\begin{aligned} & 658929 \\ & 3 \end{aligned}$ | $\begin{aligned} & 296117 \\ & 8 \end{aligned}$ | $\begin{aligned} & 333675 \\ & 8 \end{aligned}$ | $\begin{aligned} & 429660 \\ & 5 \end{aligned}$ | $\begin{aligned} & 263011 \\ & 0 \end{aligned}$ | 770070 | 0.28 | 0.35 | 0.22 |

* RCT3 estimate.
** Geometric mean 1990-2019.
$\dagger$ Estimated value from the forecast.

Table 8.7.3.2. NE Atlantic Mackerel. ESTIMATED POPULATION ABUNDANCE

| Units:Thousands |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 0 | 4811944 | 4534674 | 3958511 | 3818369 | 4135091 | 4044346 | 4002878 | 4028611 | 3736555 | 3575256 |
| 1 | 4906985 | 4565117 | 4423099 | 2843973 | 2671751 | 4185674 | 3365317 | 3309721 | 3999914 | 3008842 |
| 2 | 2352319 | 4073393 | 4218329 | 4206696 | 2001915 | 1834723 | 4127487 | 2744499 | 2686413 | 3878239 |
| 3 | 946215 | 1895555 | 3410854 | 4050384 | 4288850 | 1366283 | 1274935 | 4071426 | 2175748 | 2352879 |
| 4 | 1634417 | 727096 | 1423284 | 2873361 | 3706346 | 4055632 | 1011884 | 860882 | 3774327 | 1688475 |
| 5 | 3502369 | 1211575 | 522286 | 974609 | 2188505 | 3047018 | 3179966 | 793384 | 539031 | 3020884 |
| 6 | 2698169 | 2450353 | 867262 | 383786 | 666298 | 1626455 | 2228626 | 2173505 | 604829 | 346712 |
| 7 | 802869 | 1805822 | 1637759 | 584461 | 268795 | 462096 | 1081089 | 1497106 | 1410459 | 465834 |
| 8 | 298539 | 550334 | 1240000 | 1121849 | 396720 | 192990 | 309071 | 762959 | 1032937 | 1062503 |
| 9 | 825062 | 204624 | 376826 | 851091 | 766625 | 274128 | 135828 | 205838 | 536597 | 717372 |
| 10 | 222856 | 565887 | 140182 | 257707 | 583219 | 522820 | 191155 | 92645 | 136364 | 364659 |
| 11 | 326164 | 152766 | 387492 | 95996 | 176141 | 398065 | 354576 | 129493 | 62794 | 87873 |
| 12 | 674935 | 686985 | 574941 | 656675 | 512830 | 469173 | 586231 | 631401 | 508358 | 379329 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 3369803 | 3351837 | 3326467 | 3138516 | 3023339 | 2938787 | 3002031 | 2993152 | 3048761 | 3293974 |
| 1 | 3122075 | 2615669 | 2885102 | 3098425 | 2585116 | 2539144 | 2302095 | 2669526 | 2455580 | 2667233 |
| 2 | 2372219 | 2660273 | 1987271 | 2427002 | 2816896 | 2075235 | 2080200 | 1758428 | 2314624 | 1963068 |
| 3 | 3940417 | 2140088 | 2559852 | 1644365 | 1987870 | 2401889 | 2173801 | 1941637 | 1231726 | 2379204 |
| 4 | 1842341 | 3069095 | 1526566 | 2037733 | 1095936 | 1426580 | 1816052 | 1786099 | 1641456 | 1259088 |
| 5 | 1079949 | 1252621 | 1937084 | 988781 | 1386293 | 677896 | 971840 | 1209516 | 1522591 | 1270361 |
| 6 | 1990828 | 775860 | 949151 | 1158070 | 584868 | 973641 | 492322 | 730339 | 861227 | 903114 |
| 7 | 214963 | 1227778 | 471291 | 569476 | 649669 | 343127 | 574871 | 321440 | 481979 | 618348 |
| 8 | 352590 | 137127 | 733390 | 310132 | 339398 | 281855 | 214493 | 347588 | 264687 | 311353 |
| 9 | 722972 | 249253 | 88658 | 412282 | 183807 | 178892 | 136670 | 152721 | 212067 | 181519 |
| 10 | 464602 | 490373 | 160056 | 53134 | 216892 | 111103 | 94119 | 86468 | 103085 | 131703 |


| 11 | 242415 | 291894 | 307588 | 97913 | 30040 | 133054 | 64441 | 49787 | 53266 | 63649 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 12 | 307504 | 356962 | 413037 | 448912 | 334155 | 220092 | 214507 | 173790 | 142603 | 125449 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 0 | 3271449 | 4292777 | 4815423 | 4130904 | 4802786 | 5625640 | 5667557 | 4995749 | 4707792 | 4664141 |
| 1 | 3069524 | 1905634 | 5055227 | 6085740 | 2853647 | 3788048 | 5562698 | 5232261 | 4119724 | 3991675 |
| 2 | 2282784 | 2595111 | 1160261 | 4805824 | 6767442 | 2348077 | 3340129 | 4693422 | 4700668 | 3396575 |
| 3 | 1842612 | 1751779 | 2524613 | 792439 | 3971769 | 5359865 | 1678094 | 2442601 | 4353419 | 4927061 |
| 4 | 1844107 | 1308343 | 1555153 | 1568371 | 753796 | 1863630 | 3158145 | 1454237 | 1944926 | 3896436 |
| 5 | 1037294 | 1251170 | 994749 | 921241 | 1008506 | 536915 | 1021782 | 2077797 | 1227021 | 1588070 |
| 6 | 862156 | 678869 | 813023 | 582841 | 479413 | 477746 | 372778 | 748714 | 1106484 | 907378 |
| 7 | 619208 | 607205 | 414817 | 383143 | 268963 | 231907 | 280874 | 254789 | 421706 | 693813 |
| 8 | 375071 | 412556 | 349520 | 245409 | 186923 | 135367 | 131299 | 185038 | 178773 | 265207 |
| 9 | 190676 | 240199 | 230717 | 197698 | 118090 | 87514 | 73349 | 95048 | 102537 | 109904 |
| 10 | 113401 | 128306 | 128642 | 119191 | 93961 | 62992 | 52736 | 47701 | 59082 | 53007 |
| 11 | 69909 | 68730 | 63614 | 67505 | 47920 | 31275 | 31986 | 34478 | 22178 | 28775 |
| 12 | 122154 | 127574 | 113255 | 82479 | 57594 | 40348 | 38469 | 39885 | 31447 | 20450 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| 0 | 5334499 | 5942633 | 5531582 | 5385707 | 5476329 | 5171170 | 5760094 | 5990287 | 6041636 | 6511865 |
| 1 | 4157195 | 5347146 | 6063029 | 4400260 | 4140907 | 5638931 | 3535950 | 5147012 | 4157675 | 4085909 |
| 2 | 3868325 | 3287650 | 5496980 | 6407927 | 3633802 | 3302939 | 5043805 | 2171821 | 5012797 | 2968396 |
| 3 | 3337197 | 3625580 | 2687931 | 5237881 | 6886128 | 2894679 | 2655306 | 4399675 | 1391233 | 3595444 |
| 4 | 4653138 | 3033068 | 2956488 | 2402430 | 5041550 | 4608605 | 2702920 | 2090078 | 2795349 | 944287 |
| 5 | 2934771 | 3375706 | 2376216 | 2458937 | 2391391 | 3592391 | 3380966 | 2094232 | 1343719 | 1457244 |
| 6 | 1266128 | 2128897 | 2415490 | 2204974 | 2341222 | 1947569 | 2822270 | 2867086 | 1439873 | 1044754 |
| 7 | 567224 | 916877 | 1361364 | 1608738 | 2009982 | 1829486 | 1558096 | 2537033 | 2178762 | 1045851 |
| 8 | 379544 | 415737 | 602998 | 861848 | 1342799 | 1506580 | 1335250 | 1254062 | 1787575 | 1533512 |
| 9 | 170567 | 207787 | 269303 | 409522 | 600520 | 959816 | 891519 | 1058229 | 978854 | 1263562 |
| 10 | 74593 | 94860 | 125480 | 167550 | 268317 | 457779 | 524716 | 655479 | 664292 | 585991 |
| 11 | 25440 | 46199 | 52475 | 81814 | 91627 | 135696 | 224733 | 358678 | 461162 | 405285 |
| 12 | 32038 | 39467 | 49456 | 69104 | 65793 | 100889 | 133922 | 260203 | 331065 | 441157 |
| year |  |  |  |  |  |  |  |  |  |  |
| age 2020 |  |  |  |  |  |  |  |  |  |  |
| 06597436 |  |  |  |  |  |  |  |  |  |  |
| 15442333 |  |  |  |  |  |  |  |  |  |  |
| 22773562 |  |  |  |  |  |  |  |  |  |  |
| 32031825 |  |  |  |  |  |  |  |  |  |  |
| 42325558 |  |  |  |  |  |  |  |  |  |  |
| 5815955 |  |  |  |  |  |  |  |  |  |  |
| $6 \quad 1175515$ |  |  |  |  |  |  |  |  |  |  |
| 7987483 |  |  |  |  |  |  |  |  |  |  |
| 81021186 |  |  |  |  |  |  |  |  |  |  |
| 91197961 |  |  |  |  |  |  |  |  |  |  |
| 101075322 |  |  |  |  |  |  |  |  |  |  |
| 11 | 509948 |  |  |  |  |  |  |  |  |  |
| 12 | 715077 |  |  |  |  |  |  |  |  |  |

## Table 8.7.3.3. NE Atlantic Mackerel. ESTIMATED FISHING MORTALITY

```
1981
    0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.008
    0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031
    0.059 0.059 0.059 0.059 0.059 0.059 0.059 0.059 0.059 0.059 0.059 0.060 0.061 0.061
    0.112 0.112 0.112 0.112 0.113 0.115 0.117 0.119 0.121 0.124 0.127 0.130 0.132 0.136
    0.182 0.183 0.183 0.184 0.185 0.187 0.191 0.197 0.201 0.208 0.213 0.218 0.222 0.225
    0.207 0.207 0.208 0.210 0.211 0.214 0.218 0.222 0.227 0.232 0.237 0.242 0.251 0.257
    0.253 0.254 0.255 0.256 0.259 0.263 0.267 0.273 0.278 0.289 0.299 0.308 0.316 0.324
    0.228 0.228 0.228 0.229 0.230 0.233 0.238 0.244 0.252 0.264 0.281 0.304 0.329 0.352
    0.228 0.228 0.228 0.229 0.230 0.233 0.238 0.244 0.252 0.264 0.281 0.304 0.329 0.352
    0.228 0.228 0.228 0.229 0.230 0.233 0.238 0.244 0.252 0.264 0.281 0.304 0.329 0.352
    0 0.228 0.228 0.228 0.229 0.230 0.233 0.238 0.244 0.252 0.264 0.281 0.304 0.329 0.352
    0.228 0.228 0.228 0.229 0.230 0.233 0.238 0.244 0.252 0.264 0.281 0.304 0.329 0.352
    2 0.228 0.228 0.228 0.229 0.230 0.233 0.238 0.244 0.252 0.264 0.281 0.304 0.329 0.352
        year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007
    0.008 0.008 0.008 0.008 0.008 0.008 0.007 0.007 0.007 0.006 0.005 0.005 0.005 0.005
    0.030 0.030 0.030 0.030 0.030 0.030 0.029 0.028 0.027 0.024 0.021 0.019 0.019 0.018
    0.062 0.063 0.063 0.065 0.066 0.067 0.068 0.068 0.067 0.066 0.068 0.063 0.055 0.046
    0.138 0.140 0.143 0.146 0.149 0.156 0.164 0.158 0.158 0.144 0.146 0.135 0.115 0.107
    0.228 0.229 0.230 0.230 0.235 0.242 0.254 0.263 0.259 0.235 0.222 0.196 0.182 0.176
    0.261 0.266 0.273 0.285 0.300 0.314 0.331 0.321 0.325 0.320 0.309 0.278 0.254 0.261
    0.327 0.329 0.329 0.331 0.336 0.348 0.366 0.401 0.396 0.399 0.381 0.343 0.331 0.327
    0.367 0.363 0.346 0.333 0.335 0.347 0.357 0.406 0.464 0.510 0.467 0.363 0.341 0.411
    0.367 0.363 0.346 0.333 0.335 0.347 0.357 0.406 0.464 0.510 0.467 0.363 0.341 0.411
    0.367 0.363 0.346 0.333 0.335 0.347 0.357 0.406 0.464 0.510 0.467 0.363 0.341 0.411
    0}00.3670.363 0.346 0.333 0.335 0.347 0.357 0.406 0.464 0.510 0.467 0.363 0.341 0.411
    1 0.367 0.363 0.346 0.333 0.335 0.347 0.357 0.406 0.464 0.510 0.467 0.363 0.341 0.411
    2 0.367 0.363 0.346 0.333 0.335 0.347 0.357 0.406 0.464 0.510 0.467 0.363 0.341 0.411
        year
age 2008 2009 2010 2011 2012 2013 2014 2015 2016 2017 2018 2019 2020
```



```
    1 0.016 0.015 0.015 0.014 0.014 0.013 0.013 0.014 0.013 0.012 0.012 0.012 0.011
    2 0.041 0.039 0.039 0.039 0.040 0.041 0.042 0.043 0.045 0.046 0.049 0.047 0.044
    0.104 0.103 0.102 0.099 0.094 0.093 0.102 0.102 0.109 0.114 0.112 0.110 0.105
    0.176 0.182 0.183 0.179 0.173 0.178 0.180}0.165 0.177 0.180 0.160 0.145 0.138
    0.254 0.247 0.246 0.237 0.232 0.230}00.248 0.226 0.218 0.225 0.231 0.239 0.244
    0.303 0.298 0.282 0.276 0.262 0.253 0.269 0.263 0.236 0.228 0.243 0.254 0.270
    0.403 0.347 0.335 0.333 0.302 0.303 0.292 0.272 0.229 0.226 0.229 0.219 0.296
    0.403 0.347 0.335 0.333 0.302 0.303 0.292 0.272 0.229 0.226 0.229 0.219 0.296
    0.403 0.347 0.335 0.333 0.302 0.303 0.292 0.272 0.229 0.226 0.229 0.219 0.296
```



```
    110.403 0.347 0.335 0.333 0. 302 0.303 0.292 0.272 0.229 0.226 0.229 0.219 0.296
    2 0.403 0.347 0.335 0.333 0.302 0.303 0.292 0.272 0.229 0.226 0.229 0.219 0.296
```

Table 8.7.5.1. NE Atlantic Mackerel. Comparison of estimated SAM parameters (and uncertainty) between the 2021 WGIWDE assessment and an assessment starting at age 2.

|  | parameters values |  |  | parameter standard deviation |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | current | Age 2 | \% difference | current | Age 2 | \% difference |
| observation standard deviations |  |  |  |  |  |  |
| Catches age 0 | 0.91 | X |  | 0.18 | X |  |
| Catches age 1 | 0.36 | X |  | 0.23 | X |  |
| Catches age 2-12 | 0.11 | 0.11 | 3\% | 0.16 | 0.15 | -9\% |
| Egg survey | 0.31 | 0.32 | 1\% | 0.26 | 0.26 | 0\% |
| Recruitment index | 0.28 | X |  | 0.30 | X |  |
| IESSNS age 3 | 0.65 | 0.61 | -5\% | 0.24 | 0.24 | -1\% |
| IESSNS ages 4-11 | 0.39 | 0.39 | 2\% | 0.14 | 0.14 | 3\% |
| Recapture overdispersion tags | 4.33 | 4.25 | -2\% | 0.25 | 0.24 | -2\% |
| process variances |  |  |  |  |  |  |
| F age 0 | 0.25 | X |  | 0.49 | X |  |
| F age 1 | 0.15 | X |  | 0.49 | X |  |
| F age 2+ | 0.13 | 0.16 | 24\% | 0.19 | 0.14 | -24\% |
| Rec Var | 0.16 | 0.55 | 246\% | 0.74 | 0.15 | -80\% |
| Proc Err Var | 0.21 | 0.18 | -16\% | 0.09 | 0.10 | 14\% |
| catchabilities |  |  |  |  |  |  |
| egg survey | 1.22 | 1.23 | 1\% | 0.11 | 0.11 | -1\% |
| recruitment index | 0.00 | X |  | 0.13 | X |  |
| IESSNS age 3 | 0.82 | 0.87 | 6\% | 0.23 | 0.22 | -5\% |
| IESSNS age 4 | 1.25 | 1.26 | 1\% | 0.16 | 0.16 | -2\% |
| IESSNS age 5 | 1.71 | 1.68 | -2\% | 0.16 | 0.16 | -1\% |
| IESSNS age 6 | 1.83 | 1.79 | -2\% | 0.16 | 0.16 | -1\% |
| IESSNS age 7 | 1.95 | 1.96 | 0\% | 0.16 | 0.16 | 0\% |
| IESSNS age 8 | 1.85 | 1.86 | 1\% | 0.16 | 0.16 | 0\% |
| IESSNS age 9 | 1.95 | 1.96 | 1\% | 0.16 | 0.16 | 0\% |
| IESSNS ages 10-11 | 1.76 | 1.77 | 0\% | 0.16 | 0.16 | 0\% |


|  | parameters values |  | parameter standard <br> deviation |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | current | Age 2 | \% difference | current | Age 2 | \% difference |
| logitReleaseSurvival_0 | 0.67 | 0.64 | $-4 \%$ | 0.11 | 0.10 | $-10 \%$ |
| logitReleaseSurvival_1 | 0.17 | 0.17 | $2 \%$ | 0.11 | 0.11 | $-4 \%$ |

Table 8.8.3.1. NE Atlantic Mackerel. Short-term prediction: INPUT DATA

|  |  | $\Sigma$ | $\frac{\lambda}{\sum_{n}^{N}}$ |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2021 |  |  |  |  |  |  |  |  |
| 0 | 4367513 | 0.15 | 0.000 | 0.000 | 0.301 | 0.000 | 0.002 | 0.060 |
| 1 | 4935722 | 0.15 | 0.092 | 0.502 | 0.301 | 0.067 | 0.012 | 0.144 |
| 2 | 4631377 | 0.15 | 0.443 | 0.182 | 0.301 | 0.197 | 0.047 | 0.264 |
| 3 | 2123273 | 0.15 | 0.905 | 0.182 | 0.301 | 0.256 | 0.110 | 0.313 |
| 4 | 1673559 | 0.15 | 0.998 | 0.182 | 0.301 | 0.289 | 0.149 | 0.358 |
| 5 | 1724965 | 0.15 | 1.000 | 0.271 | 0.301 | 0.324 | 0.239 | 0.391 |
| 6 | 418933 | 0.15 | 1.000 | 0.271 | 0.301 | 0.345 | 0.256 | 0.419 |
| 7 | 948935 | 0.15 | 1.000 | 0.271 | 0.301 | 0.362 | 0.247 | 0.437 |
| 8 | 612978 | 0.15 | 1.000 | 0.271 | 0.301 | 0.377 | 0.247 | 0.453 |
| 9 | 700155 | 0.15 | 1.000 | 0.271 | 0.301 | 0.395 | 0.247 | 0.471 |
| 10 | 741590 | 0.15 | 1.000 | 0.271 | 0.301 | 0.423 | 0.247 | 0.495 |
| 11 | 696832 | 0.15 | 1.000 | 0.271 | 0.301 | 0.438 | 0.247 | 0.513 |
| 12+ | 784248 | 0.15 | 1.000 | 0.271 | 0.301 | 0.486 | 0.247 | 0.543 |
| 2022 |  |  |  |  |  |  |  |  |
| 0 | 4367513 | 0.15 | 0.000 | 0.000 | 0.301 | 0.000 | 0.002 | 0.060 |
| 1 | - | 0.15 | 0.092 | 0.502 | 0.301 | 0.067 | 0.012 | 0.144 |
| 2 | - | 0.15 | 0.443 | 0.182 | 0.301 | 0.197 | 0.047 | 0.264 |
| 3 | - | 0.15 | 0.905 | 0.182 | 0.301 | 0.256 | 0.110 | 0.313 |
| 4 | - | 0.15 | 0.998 | 0.182 | 0.301 | 0.289 | 0.149 | 0.358 |
| 5 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.324 | 0.239 | 0.391 |
| 6 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.345 | 0.256 | 0.419 |


|  |  | $\Sigma$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.362 | 0.247 | 0.437 |
| 8 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.377 | 0.247 | 0.453 |
| 9 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.395 | 0.247 | 0.471 |
| 10 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.423 | 0.247 | 0.495 |
| 11 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.438 | 0.247 | 0.513 |
| 12+ | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.486 | 0.247 | 0.543 |
| 2023 |  |  |  |  |  |  |  |  |
| 0 | 4367513 | 0.15 | 0.000 | 0.000 | 0.301 | 0.000 | 0.002 | 0.060 |
| 1 | - | 0.15 | 0.092 | 0.502 | 0.301 | 0.067 | 0.012 | 0.144 |
| 2 | - | 0.15 | 0.443 | 0.182 | 0.301 | 0.197 | 0.047 | 0.264 |
| 3 | - | 0.15 | 0.905 | 0.182 | 0.301 | 0.256 | 0.110 | 0.313 |
| 4 | - | 0.15 | 0.998 | 0.182 | 0.301 | 0.289 | 0.149 | 0.358 |
| 5 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.324 | 0.239 | 0.391 |
| 6 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.345 | 0.256 | 0.419 |
| 7 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.362 | 0.247 | 0.437 |
| 8 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.377 | 0.247 | 0.453 |
| 9 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.395 | 0.247 | 0.471 |
| 10 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.423 | 0.247 | 0.495 |
| 11 | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.438 | 0.247 | 0.513 |
| 12+ | - | 0.15 | 1.000 | 0.271 | 0.301 | 0.486 | 0.247 | 0.543 |

Table 8.8.3.2. NE Atlantic Mackerel. Short-term prediction: Multi-option table for 1199103 t catch in 2021 and a range of F-values in 2022.

| 2021 |  |  |  |
| :--- | :--- | :--- | :--- |
| TSB | SSB | Fbar | Catch |
| 4828401 | 3510849 | 0.354 | 1199103 |


| 2022 |  |  |  | 2023 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TSB | SSB | Fbar | Catch | TSB | SSB | Implied change in the catch |
| 4419995 | 3479949 | 0.00 | 0 | 4918697 | 4056937 | -100.0\% |
| - | 3473113 | 0.01 | 33966 | 4890260 | 4022211 | -97.2\% |
| - | 3466295 | 0.02 | 67639 | 4862072 | 3987876 | -94.4\% |
| - | 3459494 | 0.03 | 101023 | 4834131 | 3953927 | -91.6\% |
| - | 3452710 | 0.04 | 134118 | 4806433 | 3920358 | -88.8\% |
| - | 3445943 | 0.05 | 166930 | 4778976 | 3887167 | -86.1\% |
| - | 3439194 | 0.06 | 199460 | 4751758 | 3854347 | -83.4\% |
| - | 3432461 | 0.07 | 231710 | 4724778 | 3821894 | -80.7\% |
| - | 3425746 | 0.08 | 263685 | 4698031 | 3789803 | -78.0\% |
| - | 3419048 | 0.09 | 295386 | 4671516 | 3758070 | -75.4\% |
| - | 3412366 | 0.10 | 326816 | 4645232 | 3726691 | -72.7\% |
| - | 3405702 | 0.11 | 357978 | 4619174 | 3695661 | -70.1\% |
| - | 3399055 | 0.12 | 388874 | 4593342 | 3664975 | -67.6\% |
| - | 3392424 | 0.13 | 419507 | 4567733 | 3634630 | -65.0\% |
| - | 3385810 | 0.14 | 449879 | 4542345 | 3604621 | -62.5\% |
| - | 3379213 | 0.15 | 479994 | 4517176 | 3574945 | -60.0\% |
| - | 3372633 | 0.16 | 509853 | 4492223 | 3545596 | -57.5\% |
| - | 3366070 | 0.17 | 539459 | 4467485 | 3516571 | -55.0\% |
| - | 3359522 | 0.18 | 568814 | 4442959 | 3487866 | -52.6\% |
| - | 3352992 | 0.19 | 597921 | 4418643 | 3459476 | -50.1\% |
| - | 3346478 | 0.20 | 626783 | 4394536 | 3431399 | -47.7\% |
| - | 3339981 | 0.21 | 655400 | 4370635 | 3403630 | -45.3\% |
| - | 3333500 | 0.22 | 683777 | 4346938 | 3376165 | -43.0\% |


| 2022 |  |  |  | 2023 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TSB | SSB | Fbar | Catch | TSB | SSB | Implied change in the catch |
| - | 3327035 | 0.23 | 711915 | 4323443 | 3349000 | -40.6\% |
| - | 3320587 | 0.24 | 739817 | 4300148 | 3322132 | -38.3\% |
| - | 3314155 | 0.25 | 767485 | 4277052 | 3295558 | -36.0\% |
| - | 3307739 | 0.26 | 794920 | 4254153 | 3269273 | -33.7\% |
| - | 3301339 | 0.27 | 822126 | 4231447 | 3243274 | -31.4\% |
| - | 3294956 | 0.28 | 849105 | 4208935 | 3217558 | -29.2\% |
| - | 3288588 | 0.29 | 875858 | 4186613 | 3192120 | -27.0\% |
| - | 3282237 | 0.30 | 902388 | 4164480 | 3166958 | -24.7\% |
| - | 3275902 | 0.31 | 928697 | 4142534 | 3142068 | -22.6\% |
| - | 3269583 | 0.32 | 954787 | 4120773 | 3117447 | -20.4\% |
| - | 3263279 | 0.33 | 980661 | 4099195 | 3093092 | -18.2\% |
| - | 3256992 | 0.34 | 1006320 | 4077800 | 3068999 | -16.1\% |
| - | 3250720 | 0.35 | 1031766 | 4056584 | 3045165 | -14.0\% |
| - | 3244464 | 0.36 | 1057002 | 4035546 | 3021586 | -11.9\% |
| - | 3238224 | 0.37 | 1082029 | 4014685 | 2998261 | -9.8\% |
| - | 3232000 | 0.38 | 1106849 | 3993998 | 2975185 | -7.7\% |
| - | 3225791 | 0.39 | 1131465 | 3973485 | 2952356 | -5.6\% |
| - | 3219598 | 0.40 | 1155878 | 3953143 | 2929770 | -3.6\% |
| - | 3213421 | 0.41 | 1180091 | 3932971 | 2907425 | -1.6\% |
| - | 3207259 | 0.42 | 1204104 | 3912966 | 2885318 | 0.4\% |
| - | 3201112 | 0.43 | 1227921 | 3893129 | 2863446 | 2.4\% |
| - | 3194981 | 0.44 | 1251543 | 3873456 | 2841805 | 4.4\% |
| - | 3188866 | 0.45 | 1274971 | 3853946 | 2820394 | 6.3\% |
| - | 3182766 | 0.46 | 1298209 | 3834598 | 2799209 | 8.3\% |
| - | 3176681 | 0.47 | 1321256 | 3815410 | 2778247 | 10.2\% |
| - | 3170611 | 0.48 | 1344116 | 3796381 | 2757507 | 12.1\% |
| - | 3164557 | 0.49 | 1366790 | 3777509 | 2736984 | 14.0\% |
| - | 3158517 | 0.50 | 1389280 | 3758793 | 2716677 | 15.9\% |


| 2022 |  |  |  | 2023 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TSB | SSB | Fbar | Catch | TSB | SSB | Implied change in the catch |
| - | 3152493 | 0.51 | 1411587 | 3740231 | 2696584 | 17.7\% |
| - | 3146484 | 0.52 | 1433714 | 3721821 | 2676700 | 19.6\% |
| - | 3140491 | 0.53 | 1455662 | 3703563 | 2657024 | 21.4\% |
| - | 3134512 | 0.54 | 1477432 | 3685454 | 2637554 | 23.2\% |
| - | 3128548 | 0.55 | 1499027 | 3667494 | 2618287 | 25.0\% |
| - | 3122599 | 0.56 | 1520448 | 3649680 | 2599219 | 26.8\% |
| - | 3116665 | 0.57 | 1541696 | 3632012 | 2580350 | 28.6\% |
| - | 3110746 | 0.58 | 1562774 | 3614488 | 2561677 | 30.3\% |
| - | 3104841 | 0.59 | 1583682 | 3597107 | 2543196 | 32.1\% |
| - | 3098952 | 0.60 | 1604424 | 3579867 | 2524907 | 33.8\% |
| - | 3093077 | 0.61 | 1624999 | 3562767 | 2506806 | 35.5\% |
| - | 3087217 | 0.62 | 1645410 | 3545806 | 2488892 | 37.2\% |
| - | 3081371 | 0.63 | 1665658 | 3528982 | 2471162 | 38.9\% |
| - | 3075540 | 0.64 | 1685745 | 3512294 | 2453613 | 40.6\% |
| - | 3069724 | 0.65 | 1705672 | 3495741 | 2436245 | 42.2\% |
| - | 3063922 | 0.66 | 1725441 | 3479322 | 2419054 | 43.9\% |
| - | 3058135 | 0.67 | 1745053 | 3463035 | 2402038 | 45.5\% |
| - | 3052362 | 0.68 | 1764509 | 3446878 | 2385196 | 47.2\% |
| - | 3046603 | 0.69 | 1783812 | 3430852 | 2368525 | 48.8\% |
| - | 3040859 | 0.70 | 1802963 | 3414954 | 2352024 | 50.4\% |
| - | 3035130 | 0.71 | 1821962 | 3399183 | 2335689 | 51.9\% |
| - | 3029414 | 0.72 | 1840812 | 3383538 | 2319520 | 53.5\% |
| - | 3023713 | 0.73 | 1859513 | 3368018 | 2303514 | 55.1\% |
| - | 3018026 | 0.74 | 1878068 | 3352622 | 2287670 | 56.6\% |
| - | 3012353 | 0.75 | 1896478 | 3337349 | 2271985 | 58.2\% |
| - | 3006694 | 0.76 | 1914743 | 3322196 | 2256458 | 59.7\% |
| - | 3001050 | 0.77 | 1932866 | 3307164 | 2241086 | 61.2\% |
| - | 2995419 | 0.78 | 1950847 | 3292252 | 2225868 | 62.7\% |


| 2022 |  |  |  | 2023 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TSB | SSB | Fbar | Catch | TSB | SSB | Implied change in the catch |
| - | 2989803 | 0.79 | 1968688 | 3277457 | 2210802 | 64.2\% |
| - | 2984200 | 0.80 | 1986391 | 3262779 | 2195887 | 65.7\% |
| - | 2978611 | 0.81 | 2003955 | 3248217 | 2181120 | 67.1\% |
| - | 2973037 | 0.82 | 2021384 | 3233769 | 2166499 | 68.6\% |
| - | 2967476 | 0.83 | 2038678 | 3219436 | 2152024 | 70.0\% |
| - | 2961929 | 0.84 | 2055838 | 3205214 | 2137692 | 71.4\% |
| - | 2956395 | 0.85 | 2072865 | 3191105 | 2123501 | 72.9\% |
| - | 2950876 | 0.86 | 2089761 | 3177105 | 2109450 | 74.3\% |
| - | 2945370 | 0.87 | 2106528 | 3163215 | 2095538 | 75.7\% |
| - | 2939878 | 0.88 | 2123165 | 3149434 | 2081762 | 77.1\% |
| - | 2934399 | 0.89 | 2139675 | 3135760 | 2068122 | 78.4\% |
| - | 2928934 | 0.90 | 2156058 | 3122192 | 2054615 | 79.8\% |
| - | 2923483 | 0.91 | 2172316 | 3108730 | 2041239 | 81.2\% |
| - | 2918045 | 0.92 | 2188450 | 3095372 | 2027994 | 82.5\% |
| - | 2912621 | 0.93 | 2204460 | 3082118 | 2014878 | 83.8\% |
| - | 2907210 | 0.94 | 2220349 | 3068966 | 2001890 | 85.2\% |
| - | 2901812 | 0.95 | 2236118 | 3055916 | 1989027 | 86.5\% |
| - | 2896428 | 0.96 | 2251766 | 3042966 | 1976289 | 87.8\% |
| - | 2891057 | 0.97 | 2267296 | 3030116 | 1963673 | 89.1\% |
| - | 2885700 | 0.98 | 2282709 | 3017364 | 1951180 | 90.4\% |
| - | 2880355 | 0.99 | 2298005 | 3004711 | 1938806 | 91.6\% |
| - | 2875024 | 1.00 | 2313186 | 2992154 | 1926551 | 92.9\% |
| - | 2869706 | 1.01 | 2328252 | 2979694 | 1914413 | 94.2\% |
| - | 2864402 | 1.02 | 2343205 | 2967328 | 1902392 | 95.4\% |
| - | 2859110 | 1.03 | 2358047 | 2955057 | 1890485 | 96.7\% |
| - | 2853831 | 1.04 | 2372777 | 2942879 | 1878691 | 97.9\% |
| - | 2848566 | 1.05 | 2387396 | 2930793 | 1867010 | 99.1\% |
| - | 2843313 | 1.06 | 2401907 | 2918800 | 1855439 | 100.3\% |


| 2022 |  |  |  | 2023 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| TSB | SSB | Fbar | Catch | TSB | SSB | Implied change in the catch |
| - | 2838074 | 1.07 | 2416310 | 2906897 | 1843978 | $101.5 \%$ |
| - | 2832847 | 1.08 | 2430605 | 2895084 | 1832624 | $102.7 \%$ |
| - | 2827634 | 1.09 | 2444794 | 2883360 | 1821378 | $103.9 \%$ |

Table 8.8.3.3. NE Atlantic Mackerel. Short-term prediction: Management option table for 1199103 t catch in 2021 and a range of catch options in 2022.

| Rationale | Catch (2022) | $F_{\text {bar }}$ (2022) | SSB (2022) | SSB (2023) | \% SSB <br> change | \% catch change | \% advice change |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MSY approach: F = FMSY | 794920 | 0.26 | 3307739 | 3269273 | -1.2 | -33.7 | -6.7 |
| Catch (2022) = Zero | 0 | 0 | 3479949 | 4056937 | 16.6 | -100.0 | -100.0 |
| $\begin{aligned} & \text { Catch }(2022)=2021 \\ & \text { catch }-20 \% \end{aligned}$ | 959282 | 0.32 | 3268490 | 3113212 | -4.8 | -20.0 | 12.6 |
| $\begin{aligned} & \text { Catch }(2022)=2021 \\ & \text { catch } \end{aligned}$ | 1199103 | 0.42 | 3208545 | 2889918 | -9.9 | 0.0 | 40.7 |
| $\begin{aligned} & \text { Catch }(2022)=2021 \\ & \text { catch }+25 \% \end{aligned}$ | 1498879 | 0.55 | 3128589 | 2618418 | -16.3 | 25.0 | 75.9 |
| $\begin{aligned} & \text { Fbar }(2022)=\text { Fbar } \\ & (2021) \end{aligned}$ | 1041030 | 0.35 | 3248428 | 3036502 | -6.5 | -13.2 | 22.1 |
| $\begin{aligned} & \text { Fbar }(2022)=0.36 \\ & \text { (Fpa) } \end{aligned}$ | 1057002 | 0.36 | 3244464 | 3021586 | -6.9 | -11.9 | 24.0 |
| Fbar (2022) $=0.46$ (Flim) | 1298209 | 0.46 | 3182766 | 2799209 | -12.1 | 8.3 | 52.3 |
| SSB (2023) $=$ Blim | 2220349 | 0.94 | 2907210 | 2001890 | -31.2 | 85.4 | 160.8 |
| SSB (2023) = Bpa | 1541696 | 0.57 | 3116665 | 2580350 | -17.3 | 28.8 | 81.2 |

* SSB 2023 relative to SSB 2022.
** Catch in 2022 relative to estimated catches in 2021 (1 199 103 t). There is no internationally agreed TAC for 2021.
*** Advice value for 2022 relative to the advice value for 2021 (852 284 t).


### 8.15 Figures



Figure 8.4.2.1. NE Atlantic Mackerel. Commercial catches in 2020, quarter 1.


Figure 8.4.2.2. NE Atlantic Mackerel. Commercial catches in 2020, quarter 2.


Figure 8.4.2.3. NE Atlantic Mackerel. Commercial catches in 2020, quarter 3.


Figure 8.4.2.4. NE Atlantic Mackerel. Commercial catches in 2020, quarter 4.


Figure 8.5.2.1. NE Atlantic mackerel. Weights-at-age in the catch.


Figure 8.5.2.2. NE Atlantic mackerel. Weights-at-age in the stock.


Figure 8.5.3.1. NE Atlantic mackerel. Proportion of mature fish at age.


Figure 8.6.1.3.1. Number of samples for NSMEGS 2021; plankton samples per half ICES rectangle (left) and pelagic trawl hauls for mackerel adult samples (right).


Figure 8.6.1.3.2. Stage 1A mackerel egg production (eggs $/ \mathrm{m}^{2} /$ day) by half rectangle for NSMEGS 2021. Purple circles represent observed values, black circles represent interpolated values, and crosses represent observed zeros.


Figure 8.6.1.4.1.: Aggregated daily egg production values (stage 1 eggs/m2/day) by half ICES rectangle for all MEGS stations sampled in 2016 and 2019 for all periods. Egg production values are square root transformed. Crosses denote locations where sampling was undertaken but where no spawning was recorded. Area in yellow denotes the maximum geographical survey extent for the western and southern survey area. Stations ranked in descending order and half ICES rectangles capturing $50 \%$ of total spawning activity overlaid in blue.


Figure 8.6.1.4.2.: Mackerel stage 1 egg counts/ $\mathrm{m}^{2} /$ day survey 0321 H , for all stations sampled. The coloured squares represent the surface temperature in degrees Celsius at 5 m depth during the ichthyoplankton deployments. Red outlined area denotes stations completed as part of North Sea MEGS.


Figure 8.6.2.1. Demersal trawl survey data used to derive the abundance index of age-0 mackerel. (a) Trawl sample locations in the fourth quarter (Q4, October - November, blue dots); (b) trawl sample locations in the first quarter (Q1, January - March, light blue dots); (c) number of samples by year and quarter; and (d) depth.


Figure 8.6.2.2. Spatial distribution of mackerel juveniles at age 0 in October to March. On the left, average for cohorts from 1998-2020; and on the right, 2020 cohort. Mackerel squared catch rates by trawl haul (circle areas represent catch rates in $\mathrm{kg} / \mathrm{km} 2$ ) overlaid on modelled squared catch rates per $10 \times 10 \mathrm{~km}$ rectangle. Each rectangle is coloured according to the expected squared catch rate in percent of the highest value for that year. See Jansen et al. (2015) for details.


Figure 8.6.2.3. Index of mackerel juveniles at age 0 in October to March proxied by annual integration of square root of expected catch in demersal trawl surveys (Blue lines). See Jansen et al. (2015) for details. * Rescaled


Figure 8.6.3.1. Estimated total stock numbers (TSN) of mackerel from IESSNS calculated using StoX for the years 2007 and from 2010 to 2021. Displayed is StoX baseline estimate (red dot) and a bootstrap estimate (black dot), calculated using 1000 replicates, with $90 \%$ confidence intervals (vertical line) based on the bootstrap. Analysis excludes the North Sea and survey coverage was incomplete in 2007 and 2011.


Figure 8.6.3.2. Estimated total stock biomass of mackerel from IESSNS calculated using StoX for the years 2007 and from 2010 to 2021. Displayed is StoX baseline estimate (red dot) and a bootstrap estimate (black dot), calculated using 1000 replicates, with $90 \%$ confidence intervals (vertical line) based on the bootstrap. Analysis excludes the North Sea and survey coverage was incomplete in 2007 and 2011.


Log10 (index+1)

Figure 8.6.3.3. Internal consistency of the mackerel abundance index from the IESSNS surveys including data from 2012 to 2021, excluding North Sea. Ages indicated by white numbers in grey diagonal cells. Statistically significant positive correlations ( $p<0.05$ ) are indicated by regression lines and red cells in upper left half. Correlation coefficients ( $r$ ) are given in the lower right half.


Figure 8.6.3.4. Mackerel catch curves from the estimate stock size at age from the IESSNS in 2010 and from 2012 to 2021, excluding the North Sea. Each cohort is marked by a uniquely coloured line that connects the estimates indicated by the respective ages.


Figure 8.6.3.5. Mackerel numbers by age from the IESSNS survey in 2021, excluding North Sea. Boxplot of abundance and relative standard error (CV) obtained by bootstrapping with 500 replicates using StoX version 3.10.


Figure 8.6.3.6. Mackerel catch rates from predetermined surface trawl stations (circle size represents catch rate in $\mathrm{kg} / \mathrm{km} 2$ ) overlaid on mean catch rate per standardized rectangle ( $2^{\circ}$ lat. x $4^{\circ}$ lon.) from the 2021 IESSNS, including North Sea. Zero mackerel catches are displayed as grey crosses.


Figure 8.6.3.7. Mackerel annual distribution proxied by the absolute distribution of mean mackerel catch rates per standardized rectangles ( $2^{\circ}$ lat. x $4^{\circ}$ lon.), from predetermined surface trawl stations from IESSNS in 2010 to 2021, including North Sea. Colour scale goes from white $(=0)$ to red (= maximum value for the given year).


Figure 8.6.4.1. Number and distribution of RFID tagged mackerel from experiments west of Ireland and British Isles during 2011-2021. Note that data from releases 2011-2012 are not used in the stock assessment, based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019), and data from experiments in 2020-2021 are not included as there are no full years with recaptures yet.


Figure 8.6.4.2. Biomass and distribution of catches scanned for RFID tagged mackerel during 2012-2020. Note that data from scanned catches in 2012-2013 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019).


Figure 8.6.4.3. Distribution of recaptures of RFID tagged mackerel during 2012-2020. Note that data on recaptures in 2012-2013 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019).

8.6.4.4. Overview of the relative year class distribution among RFID tagged mackerel per release year from experiments west of Ireland and British Isles in May-June, compared with the number scanned and recaptured in year 1 and 2 after release of the same year classes. Note that data from releases in 2011-2012 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019). Note also that it was decided to only use ages 5-11 in updated assessments, and limits for this age span is marked (vertical grey dotted lines) for each release year.


Figure 8.6.4.5. Upper panel: Trends in year class abundance ( $\mathrm{N}=$ numbers released/numbers recaptured*numbers scanned) from RFID tag-recapture data based on aggregated data on recaptures and scanned numbers in year 1 and 2 after each release year. Data excluded in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019), release years 2011-2012 and ages 2-4 and 12+, are marked with dotted lines in year class trends. Bottom panel: Trends in various age aggregated biomass indices from RFID tag-recapture data compared with the SSB ( $\pm 95$ confidence intervals) from the WGWIDE2021 stock assessment. Data are based on a combination of estimated numbers by year class showed in upper panel scaled by survival parameter ( 0.1466 ) and weight at age in stock from WGWIDE2021. Vertical dotted line marks the starting year where RFID tagging experiments are used in the stock assessment. Note that final year with RFID biomass estimates in 2019 is only based on recapture year 2020 and will likely change when adding recapture year 2021 in WGWIDE2022.


Figure 8.6.4.6. Signals of total mortality rate (Z). Upper panels show the trends in abundance of year classes 2003-2014 from unscaled input data (RFID, IESSNS and catches) and the WGWIDE2021 stock assessment. The estimated slope of decrease from the age 4 when it is fully recruited to the spawning stock until age 12 is interpreted as signal Z , grey dotted lines is $Z=0.4$. Bottom panels summarize the year class differences in estimated total mortality rate (with $95 \%$ confidence intervals), and differences between the various data sources.


Figure 8.7.2.1. NE Atlantic mackerel. Parameter estimates from the SAM model (and associated confidence intervals) for the WGWIDE 2021 update assessment. top left : estimated standard deviation for the observation errors, top centre : estimated overdispersion for the errors on the tag recaptures, top right : standard deviation for the processes, bottom : survey catchabilities and post-release survival of tagged fish.


Figure 8.7.2.2. NE Atlantic mackerel. Parameter uncertainty (standard deviation of estimate) versus parameter value for the observation variances. The colours correspond to the different data sources and the number next to the dots indicate the age range to which each parameter apply.
Age 3

Figure 8.7.2.3. NE Atlantic mackerel. Estimated AR1 error correlation structure for the observations from the IESSNS survey age 3 to 11 .


Figure 8.7.2.4. NE Atlantic mackerel. Correlation between parameter estimates from the SAM model for the WGWIDE 2021 update assessment


Figure 8.7.2.5. NE Atlantic mackerel. One Step Ahead Normalized residuals for the fit to the catch data (catch data prior to 2000 in blue rectangle were not used to fit the model). Blue circles indicate positive residuals (observation larger than predicted) and filled red circles indicate negative residuals.


Figure 8.7.2.6. NE Atlantic mackerel. One step ahead residuals for the fit to the recaptures of tags in the final assessment. The $x$-axis represents the release year, and the $y$-axis is the number of years between tagging and recapture. Each panel correspond to a given age at release. Blue circles indicate positive residuals (observation larger than predicted) and filled red circles indicate negative residuals.


Figure 8.7.2.7. NE Atlantic mackerel. Leave one out assessment runs. SAM estimates of SSB, Fbar and recruitment, for assessments runs leaving out one of the observation data sets.


Figure 8.7.2.8. NE Atlantic mackerel. Estimated. Sensitivity of the estimated stock trajectories to the latest year of catch-at-age data and RFID data, and comparison with WGWIDE 2020 assessment.


Figure 8.7.3.1. NE Atlantic mackerel. Perception of the NEA mackerel stock, showing the SSB, Fbart-8 and recruitment (with $95 \%$ confidence intervals) from the SAM assessment.

## Selectivity of the Fishery by Pentad



Figure 8.7.3.2. NE Atlantic mackerel. Estimated selectivity for the period 1990 to 2021, calculated as the ratio of the estimated fishing mortality-at-age and the Fbar4-8 value in the corresponding year.


Figure 8.7.4.1. NE Atlantic mackerel. Uncertainty (standard deviation of the log values) of the estimates of SSB and Fbar from the SAM for the 2020 and 2021 WGWIDE assessments.


Figure 8.7.4.2. NE Atlantic mackerel. Analytical retrospective patterns (7 years back) of SSB, Fbar $4-8$ and recruitment from the WGWIDE 2021 update assessment. the Mohn's rho values are calculated based on 5 retro years


Figure 8.7.4.3. NE Atlantic mackerel. Process error expressed as annual deviations of abundances at age, for the 2021 WGWIDE assessment and from the 2020 WGWIDE assessment.


Figure 8.7.4.4. NE Atlantic mackerel. Model process error expressed in biomass cumulated across age-group for the 2021 WGWIDE assessment and for the 2020 WGWIDE assessment.


Figure 8.7.5.1. NE Atlantic mackerel. Model. comparison of the cohort signal based on SAM estimates at age 0 , 2 and 3.

8.7.5.2. NE Atlantic mackerel. Model. comparison of the perception of the stocks from the WGWIDE 2021 assessment, and the assessment starting at age 2 .


Figure 8.7.5.3 NE Atlantic mackerel. Model. comparison of the perception of the stocks from the WGWIDE 2021 assessment, and the assessment using the RFID data in the form of abundance index for ages 5 to 11 .


Figure 8.10.1. NE Atlantic mackerel. Comparison of the stock trajectories between the 2021 WGWIDE assessment and the 2020 WGWIDE assessment.

scaling
parameters


Figure 8.10.2. NE Atlantic mackerel. Comparison of model parameters and their uncertainty for the 2021 WGWIDE and the 2020 WGWIDE assessment


Figure 8.10.3. NE Atlantic mackerel. Comparison of the uncertainty on estimates of SSB and Fbar for the WGWIDE 2021 update assessment and the 2020 WGWIDE.


Figure 8.10.4. NE Atlantic mackerel. Comparison of the abundances at age from 2011 to 2021 estimated from the 2020 and 2021 assessments.


Figure 8.11.1. NE Atlantic mackerel. Top: comparison of the ICES advice, the agreed TAC (or the sum of the unilateral quota) and total catch. Bottom: calculated percentage of Catch over Advice (CoA) and TAC over Advice (ToA).

