

WGNAS Stock Annex for Atlantic salmon

Stock specific documentation of standard assessment procedures used by ICES.

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|----------------------|------------------------------------------------|
| Stock | Atlantic salmon (sal-nea_SA) |
| Working Group | Working Group on North Atlantic Salmon (WGNAS) |
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1. General

1.1 Stock definition

1.1.1 Background

Atlantic salmon, *Salmo salar* L., have a wide range of life-history strategies. Most forms are anadromous, however, with a juvenile phase in freshwater followed by a period at sea feeding and growing, during which the fish undergo extensive migrations in the open ocean, before they migrate back to freshwater to breed. Most Atlantic salmon return to their river of origin to spawn. This precise homing behaviour has resulted in groups of fish originating in different rivers or tributaries becoming genetically distinct as they adapt to the particular conditions that they face in their home river and along their migration routes. As a result, fish from one river or tributary can differ from fish originating in other rivers/ tributaries which have become adapted to a different set of conditions. These subgroups comprise genetically distinct 'populations' and these are regarded as the basic biological units of the Atlantic salmon species.

Large rivers and their tributaries can support several, genetically distinct populations, each with separate spawning areas within the main-stem of the river or its tributaries. In most instances, however, it is not possible to demarcate clear population boundaries within a river, and managing stocks and fisheries at this level of detail would be very complex. Thus, while there is a need to protect the sustainability of these units, the primary management unit (e.g. for reporting catch statistics and regulating fishing) is generally taken to be the river stock, comprising all fish originating in eggs laid within the river.

Atlantic salmon are native to the temperate and Subarctic regions of the North Atlantic Ocean and there are over 2000 rivers draining into the North Atlantic that support the fish, about 1500 of which discharge into the Northeast Atlantic. In this area, salmon distribution extends from northern Portugal to northern Russia and Iceland, while in the Northwest Atlantic, the species ranges from northeastern USA (Connecticut) to northern Canada (Ungava Bay).

Ideally, the management of all individual river stocks, and the fisheries that exploit them, might be based upon the status of each individual population. This is not always practical, however, particularly where decisions relate to the management of distant water salmon fisheries, which exploit large numbers of stocks originating in broad geographic areas. WGNAS has therefore had to consider how populations or river stocks should be grouped in providing management advice. For this purpose, groups have been established which fall within the meaning of a stock as 'an exploited or managed unit' (Royce, 1984) and that are consistent with the ICES (1996) definition of salmon 'stocks' as 'units of a size (encompassing one or more populations) which

provide a practical basis for the fishery manager'. The issues around the grouping of Atlantic salmon stocks for the provision of management advice are reviewed in detail in Crozier *et al.* (2003). Such stock groupings have typically been referred to as stock complexes.

Salmon mature at various sea ages, typically returning to freshwater to spawn after one to three years at sea, but also sometimes at older sea ages; this varies widely between populations. Those salmon that return after one year at sea are referred to as one-sea-winter (1SW) salmon, or grilse, with older fish categorised as 2SW, 3SW, etc. In practice, however, for management purposes these older sea age fish are typically aggregated and collectively referred to as multi-sea-winter (MSW) salmon. The sea age when salmon become sexually mature depends on genetics as well as growing conditions in the sea, and possibly freshwater, although the precise proximate factors initiating homeward migration are unknown (Hansen and Quinn, 1998). The sea age of Atlantic salmon is important in the context of stock definition since these different groups of fish have different migration routes, return at different times and are differentially exploited in fisheries. Thus, for example, it is only potential MSW salmon that are exploited in the distant water salmon fishery that operates off the west coast of Greenland.

1.1.2 Management framework for salmon in the North Atlantic

The advice generated by WGNAS is in response to Terms of Reference posed by the North Atlantic Salmon Conservation Organisation (NASCO), pursuant to its role in international management of salmon. NASCO was set up in 1984 by international convention (the Convention for the Conservation of Salmon in the North Atlantic Ocean), with a responsibility for the conservation, restoration, enhancement, and rational management of wild salmon in the North Atlantic. NASCO now has six Parties that are signatories to the Convention, Canada, Denmark (in respect of Faroe Islands and Greenland), the EU (which represents its Member States), Norway, Russia and the USA. While sovereign states retain their role in the regulation of salmon fisheries for salmon originating in their own rivers, fisheries within the jurisdiction of one Party that exploit salmon originating in the rivers of another Party may be regulated by NASCO under the terms of the Convention. This is currently the case for the distant water salmon fisheries at Greenland and Faroes.

NASCO discharges these responsibilities via three Commission areas shown below:



While homewater fisheries are not regulated directly by NASCO, national/ regional jurisdictions seek to comply with NASCO agreements and guidelines in exercising their responsibilities. In particular, NASCO's Agreement on the Adoption of a Precautionary Approach states that an objective for the management of salmon fisheries is to maintain the diversity and abundance of salmon stocks, and NASCO's Standing Committee on the Precautionary Approach interpreted this as being "to maintain both the productive capacity and diversity of salmon stocks" (NASCO, 1998).

NASCO's Action Plan for Application of the Precautionary Approach (NASCO, 1999) provides an interpretation of how this is to be achieved:

- "Management measures should be aimed at maintaining all stocks above their conservation limits by the use of management targets".
- "Socio-economic factors could be taken into account in applying the precautionary approach to fisheries management issues".
- "The precautionary approach is an integrated approach that requires, *inter alia*, that stock rebuilding programmes (including as appropriate, habitat improvements, stock enhancement, and fishery management actions) be developed for stocks that are below conservation limits".

In requesting scientific advice from ICES, NASCO asks for an annual review of events in the salmon fisheries and of the status of salmon stocks around the North Atlantic; NASCO also requests management advice for stocks in each of the Commission Areas. In fulfilling these requirements, three specific purposes have been identified for which stock groupings may be required (Crozier *et al.*, 2003):

- providing descriptions of the status of stocks;
- developing models to estimate and/or forecast pre-fishery abundance (PFA); and
- developing management advice for the distant water fisheries.

Crozier *et al.* (2003) further noted that there is no reason to assume that the same stock groupings should be used for all these purposes, indeed both the criteria used (e.g. geographical or biological features) and the resulting groups are likely to differ.

1.1.3 Stock groupings used by WGNAS in providing management advice

As noted above, Atlantic salmon would, ideally, be assessed and managed on the basis of river-specific stock units. In reality, <25% of the rivers with salmon populations in the North Atlantic are so assessed (Chaput, 2012; ICES, 2013). Consequently, stock status is often, of necessity, assessed at broader regional, national and subcontinental scales. While there might be merit in grouping stocks according to biological criteria (which could cross jurisdictional boundaries), it has generally been considered that the difficulties of collecting data in a similar format in different jurisdictions is likely to outweigh the benefits of using such groups (Crozier *et al.*, 2003). It is also recognised that compilations of data on stocks within each jurisdiction are of importance to regional / national managers. As such, regional / national stock groups are typically used by ICES in providing advice on the status of stocks, with additional information compiled on biological groups (e.g. sea ages) as required.

ICES has previously provided information on the status of stocks in the Northeast Atlantic Commission (NEAC) area by region or by country (as well as sea age). For the North American Commission (NAC) area similar information is provided for the USA and the five main provincial regions in eastern Canada: Labrador, Newfoundland, Québec, Gulf and Scotia-Fundy.

In providing management advice for the mixed-stock distant-water fisheries, broader scale stock groupings have been considered appropriate. For the NAC area this is based on the six geographic regions of North America detailed above. For the NEAC area, the following national groupings have been used in recent years to provide NASCO with catch advice or alternative management advice for the distant-water fisheries at West Greenland and Faroes.

| SOUTHERN NEAC COUNTRIES | NORTHERN NEAC COUNTRIES |
|-------------------------------------------|-------------------------------------------|
| France | Russia |
| Ireland | Finland |
| UK (Northern Ireland) | Norway |
| UK (England & Wales) | Sweden |
| UK (Scotland) | Iceland (north/east regions) ¹ |
| Iceland (south/west regions) ¹ | |

¹ The Iceland stock complex was split into two groups for stock assessment purposes in 2005 (ICES, 2005), largely on the basis of tag-recapture information. Prior to 2005, all regions of Iceland were considered part of the Northern NEAC stock complex.

These groups were deemed appropriate by WGNAS as they fulfilled an agreed set of criteria for defining stock groups for the provision of management advice that were considered in detail at the 2002 WGNAS meeting (ICES, 2002) and re-evaluated at the 2005 WGNAS meeting (ICES, 2005). ICES subsequently noted, however, that provision of catch advice for NEAC stocks in the distant water fisheries should preferably be based on a larger number of smaller management units, similar to those used in the NAC area (ICES, 2010a; 2011). Such an approach was developed at the 2013 WGNAS meeting (ICES, 2013) and indicative catch advice was provided at the country level as

well as the Southern and Northern NEAC stock complexes. ICES is awaiting feedback from NASCO on the choice of management units.

Salmon from most NEAC stocks mix in the Norwegian Sea in autumn and winter, and are exploited by the fishery at Faroes. While there is evidence that some salmon from NAC rivers have been caught in the Norwegian Sea, they are currently not considered in the NEAC assessments, although this is now under review. Recent genetic information suggests that more North American fish than previously thought were exploited in the fishery at Faroes. Further details on the results of these investigations are provided in Section 3.3.3 of this report and potential options for accounting for these fish in future catch advice is provided in Section 3.6. To date, consideration of the level of exploitation of national stocks in the Faroes fishery (ICES, 2005) resulted in the proposal that catch advice for the fishery should be based upon all NEAC area stocks and both 1SW and MSW fish.

In contrast, the fishery to the west of Greenland operates in an area where salmon from all North America and some Northeast Atlantic stocks mix in their second summer at sea. Catch advice for this fishery is thus based on non-maturing (potential MSW) fish from all regions of North America, while consideration of the level of exploitation of national stocks in the fishery from NEAC, resulted in catch advice being based upon only Southern NEAC non-maturing 1SW (potential MSW) fish (ICES, 2005).

1.2 Fisheries

Most exploitation of Atlantic salmon is restricted to fisheries close to or within the rivers of origin of the stocks; these homewater fisheries take adult fish that are mainly returning to these rivers to spawn. As noted above, these fisheries are not directly regulated by NASCO since the Parties retain responsibility for the regulation of fisheries for salmon originating in their own rivers. However, NASCO can regulate fisheries undertaken by a Party that take salmon originating in another Party's rivers, such as is the case for the distant-water fisheries at Greenland and Faroes. These fisheries take salmon originating in a large number of rivers over a wide geographical range.

1.2.1 The Northern Norwegian Sea Fishery

A longline fishery for salmon in parts of the Norwegian Sea, north of latitude 67°N, commenced in the early 1960s. Several countries participated in this fishery and the pattern of fishing, area of operation and catches changed markedly over the years. At its peak in 1970 this fishery harvested almost 1000 tonnes of salmon.

The Convention for the Conservation of Salmon in the North Atlantic Ocean, which resulted in the formation of NASCO, came into force in October 1983. The Convention created a large protected zone, free of targeted fisheries for Atlantic salmon in most areas beyond 12 nautical miles from the coast. An immediate effect was the cessation of the salmon fishery in the Northern Norwegian Sea outside the Faroes EEZ, with the last catches in this area reported in 1984 (ICES, 2013).

In the late 1980s and early 1990s, NASCO acted through diplomatic initiatives to address fishing for salmon in international waters by vessels registered to non-NASCO Parties. There have been no reports of such activities since.

1.2.2 The Faroes fishery

The fishery in the Faroes area commenced in 1968 with a small number of vessels fishing up to 70 miles north of the Faroes; initially catches increased slowly up to 40

tonnes in 1977. Danish vessels participated in the fishery between 1978 and 1982 and, at the same time, catches started to increase rapidly, peaking at 1025 tonnes in 1981. Several factors contributed to this increase: the season was extended, more vessels entered the fishery, and the fishery shifted northwards.

From 1982, the Faroese Government agreed to a voluntary quota system, involving a total catch of 750 tonnes in 1982 and 625 tonnes in 1983 (255 boats allowed 25 tonnes each). Since NASCO's establishment, regulatory measures or decisions have been agreed by the Northeast Atlantic Commission in most years (Table 1.2.2.1). These have resulted in greatly reduced allowable catches in the Faroese fishery, reflecting declining abundance of the salmon stocks. There has been no commercial salmon fishery targeting salmon around the Faroes since the early 1990s. Catches in the fishery are presented in Figure 1.2.2.1.

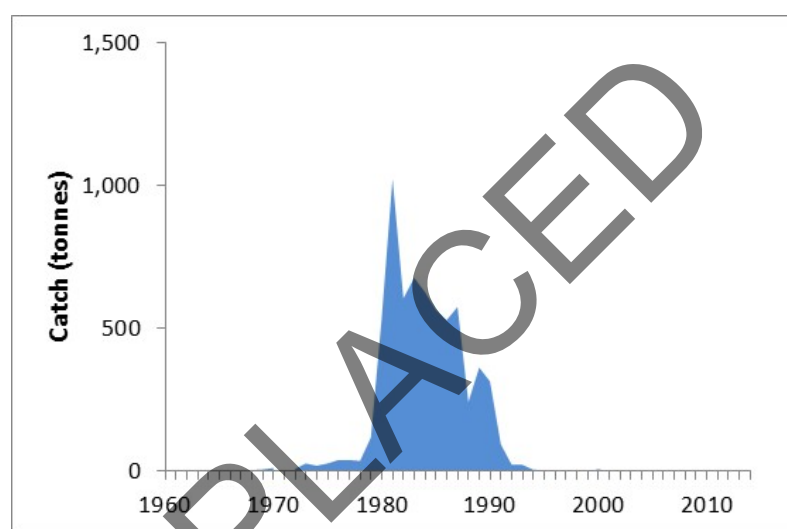


Figure 1.2.2.1. Nominal catch of salmon (tonnes, round fresh weight) in the Faroese longline fishery, 1960–2014.

Table 1.2.2.1. Summary of Regulatory Measures agreed by NASCO for the Faroese Salmon Fishery (courtesy of NASCO).

| YEAR | ALLOWABLE CATCH (TONNES) | COMMENTS/OTHER DETAILS IN THE MEASURES/DECISIONS |
|---------------|--------------------------------|--------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| 1984– 1985 | 625 | |
| 1986 | - | |
| 1987– 1989 | 1790 | Catch in any year not to exceed annual average (597t) by more than 5%. |
| 1990– 1991 | 1100 | Catch in any year not to exceed annual average (550t) by more than 15%. |
| 1992 | 550 | |
| 1993 | 550 | |
| 1994 | 550 | |
| 1995 | 550 | |
| 1996 | 470 | No more than 390 tonnes of the quota to be allocated if fishing licences issued. |
| 1997 | 425 | No more than 360 tonnes of the quota to be allocated if fishing licences issued. |
| 1998 | 380 | No more than 330 tonnes of the quota to be allocated if fishing licences issued. |
| 1999 | 330 | No more than 290 tonnes of the quota to be allocated if fishing licences issued. |
| 2000 | 300 | No more than 260 tonnes of the quota to be allocated if fishing licences issued. |
| 2001– 2003 | No quota set | It is the intention of the Faroese authorities to manage the fishery in a precautionary manner with a view to sustainability, and to make management decisions with due consideration to the advice from ICES concerning status of stocks contributing to the fishery. |
| 2004– 2006 | No quota set | It is the intention of the Faroese authorities to manage the fishery on the basis of the advice from ICES concerning status of stocks contributing to the fishery in a precautionary manner with a view to sustainability and taking into account relevant factors such as socio-economic needs and other fisheries on mixed-stocks. |
| 2007– 2012 | No quota set | It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs. |
| 2013– 2015 | No quota set | It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs. |

Note: The quotas for the Faroe Islands detailed above for the period 1984–2000 were agreed as part of effort limitation programmes (limiting the number of licences, season length and maximum number of boat fishing days) together with measures to minimise the capture of fish less than 60 cm in length. The measure for 1984/1985 did not set limits on the number of licences or the number of boat fishing days.

The Faroes salmon fishery operated over winter from November through to May. The salmon caught in the fishery originated almost entirely from European countries with salmon from many countries being present in the area (Jacobsen *et al.*, 2001). Small numbers of tagged fish originating in North America were also recaptured in the fishery (e.g. ICES, 1991), but excluded from catch advice. Recent genetic investigations, based on salmon scales removed from fish caught in the fishery in the 1980s and 1990s, suggested North American fish may make a larger contribution to the Faroes fishery than originally indicated (ICES, 2014). Genetic investigations are ongoing to confirm this and WGNAS has been asked to consider the implications of the findings in providing future catch advice to NASCO.

The fishery exploited mainly 2SW fish, although some 1 and 3SW fish were also caught. Small salmon (<60 cm total length) in their first winter at sea were required to be discarded. Large numbers of farmed salmon were also observed at Faroes and there is evidence that farmed salmon escaping from net pens in Norway entered this area (Hansen *et al.*, 1987; Hansen and Jacobsen, 2003). Such farmed fish accounted for a significant proportion of the catch; in the early 1990s, the proportion of farmed fish in this area was estimated at between 25 and 40% (Hansen *et al.*, 1999).

Tagging studies (of adult fish caught in the fishery) have indicated that some fish caught at Faroes were apparently on their way westwards, as they were reported from West Greenland later the same year (Jákupsstovu, 1988). However, salmon tagged at West Greenland were also reported in the area north of the Faroes the following year (ICES, 1984). Thus, salmon of European origin are believed to move through the Faroese area on their way to the feeding areas in the West Atlantic as well as on their return to homewaters.

1.2.3 The Greenland fishery

Limited fishing at West Greenland is reported as far back as the early 1900s, although the present fishery dates from 1959 when local fishermen began setting fixed gillnets from small boats in certain fjords around Maniitsoq (Shearer, 1992). Rapid expansion along the coast followed and from the mid-1960s Faroese and Norwegian fishermen introduced offshore driftnets, followed soon by fishermen from Greenland and Denmark. At around the same time improvements in gear (the introduction of light monofilament nets) enabled fishing in daylight and improved the efficiency of the gear. As a consequence, catches rose quickly reaching a peak of almost 2700 tonnes in 1971. Fishing by non-Greenlandic vessels was phased out in 1972–1975. However, the total catch remained at around 2000 tonnes until 1976 when a TAC of 1190 tonnes was set; the fishery has been regulated since this time. Small catches of salmon are also made on the east coast of Greenland although these are sporadic and restricted by the small number of communities in this area and by drifting polar ice.

Regulatory measures have been agreed by the West Greenland Commission for most of the years since NASCO's establishment (Table 1.2.3.1). These have resulted in greatly reduced allowable catches in the West Greenland fishery, reflecting declining abundance of the contributing salmon stocks. In all but two years since 1998, the fishery has been restricted to an internal-use fishery and commercial export of salmon is not permitted. Catches in the Greenland fishery are presented in Figure 1.2.3.1.

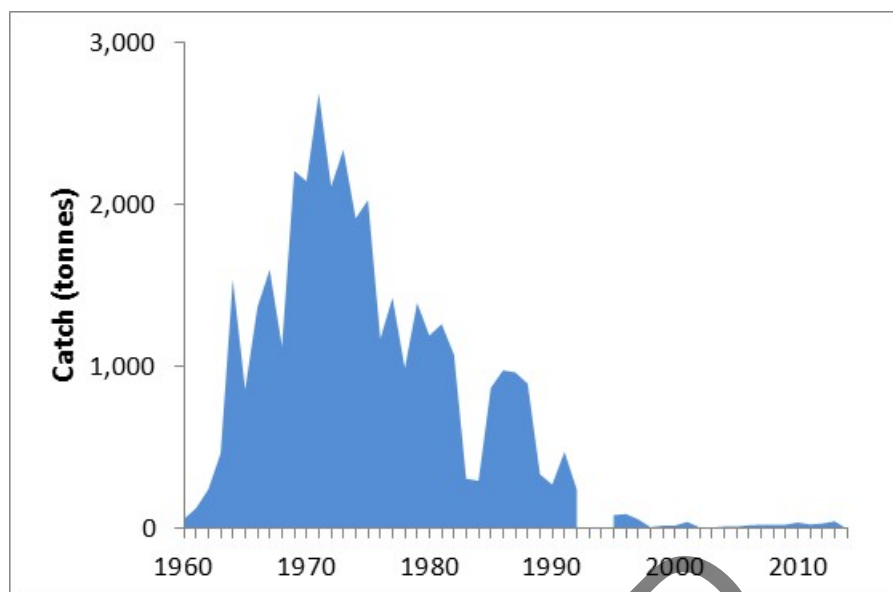


Figure 1.2.3.1. Nominal catch of salmon (tonnes, round fresh weight) in the Greenland salmon fishery, 1960–2013.

The Greenland salmon fishery operates in summer, with a fairly large proportion of the catch commonly being taken in the weeks after the opening of the season in August. Both drift and fixed gillnets continue to operate. The salmon caught in the fishery to the west of Greenland originate in both North America and the Northeast Atlantic. Data on continent of origin in the catch indicate a reasonably even split between fish from North America and Europe in the early 1990s (ICES, 2013). However, the proportion of North American fish in the catch has increased steadily since this time with North American fish comprising 80–90% of the fish caught in recent years.

The salmon caught at West Greenland are almost exclusively fish in their second summer at sea, however, these are non-maturing 1SW salmon destined to return to homewaters as 2SW, or older, fish. Fish from all parts of North America are taken in the fishery, while it is primarily only potential MSW salmon from southern countries in Europe (UK, Ireland and France) that are exploited here. Very few salmon of farmed origin appear in the catches at Greenland, and these are not taken into account in assessments.

Table 1.2.3.1. Summary of Regulatory Measures agreed by NASCO for the West Greenland Salmon Fishery (courtesy of NASCO).

| YEAR | ALLOWABLE CATCH (TONNES) | COMMENTS/OTHER DETAILS IN THE MEASURES |
|---------------|-----------------------------------------|-----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| 1984 | 870 | |
| 1985 | - | Greenlandic authorities unilaterally established quota of 852t. |
| 1986 | 850 | Catch limit adjusted for season commencing after 1 August. |
| 1987 | 850 | Catch limit adjusted for season commencing after 1 August. |
| 1988– 1990 | 2520 | Annual catch in any year not to exceed annual average (840t) by more than 10%. Catch limit adjusted for season commencing after 1 August. |
| 1991 | - | Greenlandic authorities unilaterally established quota of 840t. |
| 1992 | - | No TAC imposed by Greenlandic authorities but if the catch in first 14 days of the season had been higher compared to the previous year a TAC would have been imposed. |
| 1993 | 213 | An agreement detailing a mechanism for establishing annual quota in each of the years 1993 to 1997 was adopted by the Commission. |
| 1994 | 159 | |
| 1995 | 77 | |
| 1996 | - | Greenlandic authorities unilaterally established a quota of 174t. |
| 1997 | 57 | An addendum to the 1993 Agreement was agreed by the Commission. |
| 1998 | Internal consumption fishery only | Amount for internal consumption in Greenland has been estimated in the past to be 20 t. |
| 1999 | Internal consumption fishery only | Amount for internal consumption in Greenland has been estimated in the past to be 20 t. |
| 2000 | Internal consumption fishery only | Amount for internal consumption in Greenland has been estimated in the past to be 20 t. A Resolution Regarding the Fishing of Salmon at West Greenland was agreed by the Commission. |
| 2001 | 28 – 200 | Under an ad hoc management programme the allowable catch will be determined on the basis of cpue data obtained during the fishery. |
| 2002 | 20 - 55 | Under an <i>ad hoc</i> management programme the allowable catch will be determined on the basis of cpue data obtained during the fishery. |
| 2003– 2008 | Internal consumption fishery only | Amount for internal consumption in Greenland has been estimated in the past to be 20 t. |
| 2009– 2011 | Internal consumption fishery only | Amount for internal consumption in Greenland has been estimated in the past to be 20 t. |
| 2012– 2014 | Internal consumption fishery only | Amount for internal consumption in Greenland has been estimated in the past to be 20 t. |

1.3 Ecosystem aspects

Over the past 20 to 30 years, there has been a marked decline in the abundance of Atlantic salmon across the species' distributional range. Wild Atlantic salmon populations are declining across most of their home range and, in some cases, disappearing (ICES, 2008). Generally, populations on the southern edge of the distribution seem to have suffered the greatest decline (Parrish *et al.*, 1998; Jonsson and Jonsson, 2009; Vøllestad *et al.*, 2009). This may be linked to climatic factors. The decline in salmon abundance has coincided with a variety of environmental changes linked to an increase in greenhouse gases and a corresponding increase in temperatures (IPCC, 2001), which is most likely to have manifest effects at the edge of the species range. However, these areas are often also the ones with higher human population densities and therefore, typically, where potential impacts on the freshwater environment may also be greater. A range of factors in freshwater are known to impact on stocks including, for example, contaminants, river obstructions, and changing river flows and temperatures (ICES, 2009b; 2010b; Russell *et al.*, 2012). Such factors have potential implications for the survival of juvenile salmon and their resulting fitness when they migrate to sea as smolts (e.g. Fairchild *et al.*, 2002).

Atlantic salmon occupy three aquatic habitats during their life cycle: freshwater, estuarine and marine. Similar factors contribute to mortality in each of these habitats - competition, predation and environmental factors - but despite occurring in different habitats, these are not independent. Conditions experienced within the freshwater environment can affect the survival of emigrating smolts and marine conditions may subsequently modify the spawning success of fish in freshwater.

The decline in salmon populations has occurred despite significant reductions in exploitation, although this does not preclude possible fishery effects. An underlying cause has been a marked increase in the natural mortality of salmon at sea; the proportion of fish surviving between the smolts' seaward migration and their return to freshwater as adult fish (e.g. Peyronnet *et al.*, 2008; Chaput, 2012). For many stocks, return rates are now at the lowest levels in the time-series, even after the closure of marine fisheries. This reduced survival is thought to reflect climatic factors and broad-scale changes in ocean ecosystems as well as factors in freshwater. The exact processes controlling marine survival are relatively poorly understood (Friedland, 1998), although there is growing support for the hypotheses that survival and recruitment is mediated by growth during the post-smolt year, for European stocks at least (Friedland *et al.*, 2009).

Although their habitats are widely separated geographically, there is strong coherence in recruitment patterns between North American and European stock complexes. Recent research suggests recruitment is correlated with ocean temperature variation associated with the Atlantic Multidecadal Oscillation (AMO) (Friedland *et al.*, 2013). It further appears that there are differences in the mechanisms affecting stocks in the Northwest and Northeast Atlantic, with ocean climate variability during the first spring months of post-smolt life most important to the survival of North American stocks, while summer climate variation appears to be more important to adult recruitment variation for European stocks (Friedland *et al.*, 2013). It has been speculated that this may be related to the varying roles of predation pressure and size-related mortality on the two continental stock complexes.

In addition to changes in climate and potential issues operating in both freshwater and marine environments, various other factors have been postulated as possibly contributing to the decline in stock abundance, including predation, aquaculture

impacts and the effects of fisheries. Huge increases in aquaculture production of Atlantic salmon over recent decades (see Section 2.2.1 of the WGNAS report) have created some concerns for wild populations. The main potential impacts include: (i) genetic impacts on wild fish; (ii) discharge of organic material and other wastes; (iii) transmission of diseases and parasites (particularly sea lice) to wild populations; and (iv) concerns about obtaining adequate feed resources from an already heavily exploited marine ecosystem. For example, recent investigations in Norway have demonstrated that gene pools of wild salmon populations in a number of rivers have been gradually changed through introgression of genetic material from escaped farmed salmon (Glover *et al.*, 2012; Glover *et al.*, 2013). Sea lice also continue to be regarded as a serious problem for wild salmonids (Skilbrei *et al.*, 2013; Krkošek *et al.*, 2013) affecting their survival and perhaps also their life-history characteristics (Vollset *et al.*, 2014).

As well as declines in abundance, changes in salmon life histories are also widely reported throughout the geographic range of the species, affecting factors such as sea age composition, size at age, age-at-maturity, condition, sex ratio and growth rates (e.g. Nicieza and Braña, 1993; Hutchings and Jones, 1998; Niemelä *et al.*, 2006; Peyronnet *et al.*, 2007; Aprahamian *et al.*, 2008; Todd *et al.*, 2008). Changes are also manifest in freshwater stages, affecting factors such as the size and growth of parr and the age of smolting (e.g. Davidson and Hazelwood, 2005; Jutila *et al.*, 2006) and run timing (Kennedy and Crozier, 2010; Otero *et al.*, 2013).

2. Data

2.1 Introduction

Assessment of Atlantic salmon differs from the approaches commonly adopted for other species, for example in respect of the need for at sea surveys and collection of commercial catch per unit of effort (cpue) data. Instead, the assessment of salmon is based mainly on data collected on individual river stocks (e.g. catches and counts of returning fish), which are raised and aggregated to provide estimates of the number of fish returning to homewaters for different stock groupings. These estimates are used, in turn, to estimate abundance at earlier points in the life cycle of the fish and to inform the development of catch advice.

The provision of management advice for the mixed-stock fisheries at Faroes and West Greenland is based on assessments of the status of stocks at broad geographic scales. The North American Commission (NAC) area is divided into six management units, and the Northeast Atlantic (NEAC) Commission area is divided into 19 regions. Assessment of the status of the stocks in these areas is based on estimates of the total abundance - the pre-fishery abundance (PFA) - of different cohorts of salmon at a stage before the distant water fisheries operate. PFA is defined as the cohorts of salmon maturing as 1SW and MSW fish that are alive prior to all the marine fisheries for 1SW salmon (Rago *et al.*, 1993a). The catch advice for the NEAC area is then provided for the northern (N-NEAC) and southern (S-NEAC) stock complexes and for countries.

The models to estimate the PFA of salmon from different areas are typically based on the catch in numbers of one-sea-winter (1SW) and multi-sea-winter (MSW) salmon in each country or region, which are then raised to take account of estimates of non-reported catches and exploitation rates on the two age groups. In some cases, particularly in the NAC area, returns to homewater are estimated by alternative methods, such as counts at fishways and counting fences, or from mark and recapture studies. The estimates of fish numbers returning to homewaters are then raised to take

account of the natural mortality (M) between the date that the fish are deemed to recruit to the particular fishery of interest and the midpoint of the timing of the respective national fisheries. A value of 0.03 per month is assumed for M (Section 3.2.3). The date of recruitment of NAC stocks (and thus the PFA date) is taken as August 1st in the second summer at sea because these fish are first exploited in the distant water fishery at West Greenland. However, NEAC stocks recruit to the Faroes fishery during their first sea winter and so PFA is calculated at January 1st (i.e. eight months earlier) for these stocks.

2.2 Input data for assessments–NEAC area

PFA for NEAC stocks is estimated using the run-reconstruction approach described by Potter *et al.* (2004). The model estimates the PFA of both maturing and non-maturing 1SW salmon because both stock components may be caught in the Faroes fishery, and data for both the Faroes and West Greenland fisheries are incorporated into the model.

In order to run the NEAC PFA model, each country provides time-series (beginning in 1971) of catch in numbers, non-reporting rates and exploitation rates for 1SW and MSW salmon. Best estimates and a measure of the uncertainty or error are provided for the non-reporting and exploitation rate data in order to obtain a measure of the uncertainty in the PFA estimates, since these data are commonly derived from expert opinion. The latest data input variables used in running the NEAC assessment are listed at Appendix 3.

In some instances, the above information has been supplied in two or more regional blocks per country. In these instances, the model output is provided for the regions and is also combined to provide one set of output data for the country as a whole.

The input data for Finland consists solely of catches from the River Tana/Teno. These comprise both Finnish and Norwegian net and rod catches, as the river marks the border between these countries. The Norwegian catches from the river are not included in the input data for Norway.

Where possible, when the input data are themselves derived from other data sources, the raw data are included in the model. This allows the uncertainty in these analyses to be incorporated into the modelling approach. Thus, the catch and sample data used to estimate the catches of Scottish fish in the northeast English coastal fishery are incorporated into the assessments for both UK (England and Wales) and UK (Scotland). For Greenland, catch data are input in the form of harvests (reported and unreported) in weight, along with data from the West Greenland sampling programme.

Descriptions of how the model input data have been derived are presented below for different countries (updated from Crozier *et al.*, 2003; ICES, 2002). The methods used to derive the PFA input data for NEAC countries and options for improving the data are also discussed in Crozier *et al.* (2003).

2.2.1 Median dates of return to homewater fisheries

NEAC stocks recruit to the Faroes fishery during their first sea winter and so the date of recruitment (and thus the PFA date) is calculated at January 1st. In deriving PFA from the estimates of fish numbers returning to homewaters, it is necessary to take account of natural mortality between the date that the fish recruit to the particular fishery of interest and the midpoint of the timing of the respective national fisheries. The median return date for 1SW and MSW fish for each country/region are provided

in the table below. Thus there is about a six to nine month period between the PFA date and the median time of return to homewaters for maturing 1SW fish and 17 to 20 months for non-maturing fish.

Table 2.2.1.1. Midpoint of recruitment to homewater fisheries for NEAC countries/regions.

| NEAC COUNTRY / REGION | 1SW | MSW |
|------------------------------|-----|-----|
| Northern NEAC | | |
| Russia - Pechora River | 8 | 8 |
| Russia - Archangel / Karelia | 7.5 | 8 |
| Russia - Kola / White Sea | 8.5 | 7.5 |
| Russia - Kola / Barents Sea | 7 | 6.5 |
| Finland | 6.5 | 6 |
| Iceland - north & east | 7 | 6.5 |
| Norway | 8 | 5 |
| Sweden | 8.5 | 6 |
| Southern NEAC | | |
| Iceland - south & west | 6.5 | 6 |
| UK (Scotland - east) | 7 | 5.5 |
| UK (Scotland - west) | 8 | 5 |
| UK (N. Ireland - Fo area) | 7 | 5.5 |
| UK (N. Ireland - FB area) | 6.5 | 6 |
| Ireland | 8 | 5 |
| UK (England & Wales) | 8 | 5 |
| France | 8 | 5 |

2.2.2 Data inputs for Northern NEAC countries

2.2.2.1 Finland

Catch: The catch input to the model of Finland represents an estimate based on catch enquiries and the total number of licences issued. The Norwegian catch from the River Teno has been included in the Finnish catch, which results in a set of input data that effectively represents a single river system. Catch composition is estimated based on catch samples and corresponding scale analyses.

Level of unreported catch: Unreported catch is estimated by extrapolating the catches of the fishermen that failed to report their catches, as reporting is not mandatory.

Exploitation rates: Exploitation rates in the river fisheries are derived from radio tagging studies in 1992-93 and 1995, when 70–100 adult fish (1SW and MSW) were tagged yearly in the estuary. Most of the important river fisheries were covered by these experiments.

2.2.2.2 Norway

Area split: Salmon catches in Norway are split into four regions on the basis of climatic differences and oceanographic differences among the areas. The areas are: (1) southeast Norway from the Swedish border to the border between Rogaland and Hordaland counties, (2) southwest Norway from the border between Rogaland and Hordaland

counties to Stad (3) mid Norway from Stad to Lofoten, and (4) north Norway from Lofoten to the border with Russia.

Catch: Nominal catches of salmon in the four regions are used. In recent years there have been improvements in declaring catches. From 1979 there was a weight split 1SW/MSW (<3 kg/>3 kg). From 1993 the split was changed to 1SW/2SW/3SW (<3 kg/3–7 kg/>7 kg). Mean weight was provided for most groups and used to estimate numbers in the early part of the time-series, but in recent years the reported nominal catch (reported number of killed salmon in river and sea fisheries summed) is being used. In the input to the PFA model salmon smaller than 3 kg are regarded as 1 SW fish, whereas salmon larger than 3 kg are regarded as MSW fish. The two largest size groups are thus summed into MSW salmon. In the PFA model input the Norwegian catch data for the River Teno have been removed from the Norwegian catches and incorporated in the Finnish catches.

Unreported catch: No systematic effort is used to estimate unreported catches. Inputs are guesstimates based on occasional reports from test fishing, surveillance reports, and questionnaires. Currently there is no evidence that the level of unreported catches differs between the four regions. These estimates are provided by the management authorities.

Exploitation rates: The rates for the national model are guesstimates. For parts of southeast and southwest Norway they are derived from estimated marine exploitation rates from the River Imsa and the River Drammen. In recent years (from 2009 onwards) exploitation rates for many rivers (>50) have been taken into consideration. These exploitation rates have been obtained using a multitude of methods, mainly from drift counts of spawners or results from counting facilities combined with reported catches in the rivers. The exploitation rates have been adjusted in relation to reduced fishing effort. At present different exploitation levels are used for the different regions, reflecting different harvest regimes in the regions.

For Norway, only data from 1983 onwards have been used for assessment purposes.

2.2.2.3 Russia

Area split: The Atlantic salmon rivers of northwest Russia are split into the following four regions: Kola Peninsula - Barents Sea basin; Kola Peninsula - White Sea basin; Archangelsk Region and the Karelia; and the Pechora River region. The split is based on four regions with separate catch statistics and different biological characteristics of the stocks. For example, the difference in age composition and relative abundance of summer and autumn salmon evident among these four regions has influenced the split.

Catch: The declared catch data, in numbers, is available for the full time period (1971 onwards) for all four regions. Catches were allocated to 1SW or MSW age groups on the basis of commercial and scientific catch sampling programmes.

Level of unreported catch: Unreported catches in legal fisheries are estimated from logbooks and catch statistics, by comparing catch survey results with reported catch. Illegal catch is guesstimated and based on local knowledge of fisheries. The major component of the illegal catch in the Barents Sea basin (Kola Peninsula and Pechora River) comes from in-river fisheries and a considerable part of the illegal catch in the White Sea basin (Kola Peninsula and Archangelsk region) comes from coastal areas and this contributes the greatest uncertainties. There is a particular problem with illegal catches on the Pechora River where scientific sampling programmes suggest that the illegal catch on this river is very high. The level of non-reporting increased

considerably in the early 1990s due to the economic changes in Russia and temporary reduction of control and enforcement. Since late 2000s the higher level of non-reporting occurred in recreational fisheries due to unclear legislation for reporting. All these factors have been considered in deriving the level of unreported catch for the PFA model.

Exploitation rates: Information on exploitation rates is derived from several fisheries in the Kola Peninsula where counting fences are operated and from mark-recapture exercises on the rivers with recreational fisheries. Exploitation rates in Archangelsk and Pechora are guesstimated. These are the basis of the inputs to the model, regional sea age differences being adjusted on the basis of local knowledge from estimated stock levels.

2.2.2.4 Sweden

Catch: The catch input to the model is based on annual reported commercial salmon catch on the Swedish west coast, and on voluntary reporting from sport fishing in rivers. This reporting is detailed and considered accurate and is handled by the government agency "Swedish Agency for Marine and Water Management" (commercial catches) and the Swedish University of Agricultural Sciences (non-commercial catches). Unfortunately, reporting of catches from non-commercial fishing for salmon with gillnets or rod and line on the coast is lacking. However, due to fishing regulations these catches are small (permits required for trapnets, ban on gillnets in deeper waters, restrictions on the use of gillnets in shallow waters, limited fishing season, large marine protected areas, ban on selling fish, etc.).

There is a large proportion (ca. 60–70%) of reared fish in catches and stocks as a result of compensatory releases of reared smolts (ranching). As all ranched salmon are finclipped the catches of reared fish can be treated separately in the catch statistics. In the reporting from the commercial fishing the catch is not separated into wild and reared fish. The proportion of wild salmon is instead estimated from catch statistics in nearby rivers. Stocking of reared salmon is done in three rivers; all of these also have wild stocks in tributaries.

Catch-and-release is practised in most rivers (only rod-and-line fishing allowed in rivers) but the extent of C&R is not always known. Reported catch is landed fish, but in most rivers a proportion of the fish is released back alive and any subsequent mortality is not accounted for.

Level of unreported catch: Unreported catch, i.e. non-commercial catch of salmon in the coastal area with gillnets and rod and line, is estimated from guesstimates based on expert judgement from regional fishery officers and the Swedish University of Agricultural Sciences. These estimates are supported with catch inventories carried out in 1999 (Thörnqvist, unpubl.), 2004 (Swedish Agency for Marine and Water Management), 2008 (Thörnqvist, unpubl.). Generally, the unreported catch is estimated to be 5–10% of the reported catch.

Exploitation rates: Few fish counters are present and tagging data exist mainly for reared stocks, where the fishing pressure is higher than for wild stocks. Input for the PFA model is based on guesstimates. In the index River Ätran, data on size and composition of the spawning run and estimates of exploitation are developed at present. Since 2000, a fish ladder with an automatic counter has provided data on the spawning run in this river. Counter data in combination with results from small-scale tagging in this river are used to provide estimates of exploitation rates. One problem is that exploitation rates differ considerably between rivers. During the period 2000–

2014 the average exploitation rates for the Swedish stock as a whole have been estimated at 34% for 1SW and 39% for MSW. The exploitation rate has increased in 2011-2014 due to increased gillnet fishing on the coast. This fishery has been closed and exploitation is expected to decrease in future.

2.2.2.5 Iceland

Area split: The input data for the PFA model is divided into two areas. Rivers in the west and south of Iceland are combined into one area and rivers in the north and east into another. This is done on the basis of historic tag recoveries in ocean fisheries (which occurred in different areas) and different climate and oceanic conditions affecting the salmon life cycle, e.g. run-timing, smolt age, and sea age. The southern and western parts of Iceland fall within the NEAC southern area, while the northern and eastern parts of Iceland fall within the NEAC northern area.

Catch: Age-class information is available from individual catch records from logbooks used in the rod fishery. The division into sea age classes is based on a bimodal weight distribution. The 1SW females are <3.5 kg and 2SW females >3.5 kg, while 1SW males are <4 kg and 2SW >4 kg. Scale analyses have shown that the presence of salmon having spent more than two winters at sea and of previous spawners is uncommon and that the categorisation into 1SW and 2SW age classes by weight is regarded as fairly accurate. The net catches are recorded on a daily basis. The age split in the net fishery is derived from the weight distribution in the rod fishery from the same river system or from rivers in the same area.

In the River Ranga in southern Iceland substantial smolt releases have occurred since the early 1990s and have now reached a level of 300 000 to 500 000 smolts annually. Originally, the River Ranga had a small salmon stock with an annual catch of 10 to 90 fish until 1990. The river has very limited habitat for salmon production, but these 'ranched' fish now support a substantial rod fishery. The catch in the River Ranga comprised 23% (18–27%) of the total reported salmon rod catch in Iceland between 2009 and 2013. Since these fish are expected to have very low spawning success in the river they are excluded from the PFA catch input data.

Level of unreported catch: The fishing rights in Icelandic salmon rivers belong to landowners who must, by law, form a fishery association to manage the fishing right. The rod fishing rights are leased to the highest bidder. No ocean or estuary fisheries are allowed. The unreported catch was originally believed to be low with a guesstimate value of 2% applied. With increased use of midwater trawls in pelagic fisheries off the coast of Iceland, new information was provided which suggested an increased level of salmon bycatch. Based on a questionnaire survey, the value of unreported catch was therefore revised after 1995 to a value of 10% of the declared salmon catch. However, more recent analyses of DNA, as well as scale analyses, from salmon sampled as bycatch by Icelandic fishing vessels, indicates a low percentage of Icelandic salmon. Based on this, and other available information, a new estimate of unreported catch is now applied for Iceland at 4% of the declared catch for 1SW and MSW salmon.

Exploitation rates: Rates of rod exploitation are based on rivers with fish counters and catch records from logbooks. The estimates of exploitation are 40–50% for 1SW salmon and 50% to over 70% for 2SW salmon. The exploitation estimate for an in-river gillnet fishery is 39% to 52%, with a higher exploitation rate on larger fish. Information on the number of fish subject to catch and release in rod fisheries is also available from logbooks. The proportion of released fish has been increasing since 1996. The reduced

exploitation due to catch and release is taken into account in the annual estimate of exploitation for both 1SW and 2SW stock components in the PFA model inputs.

Median return date of 1SW and MSW: Run timing can vary both between years and between areas. The median return date of 1SW and 2SW salmon in south and west Iceland is mid-June and early June respectively. The median date of return is later in the north and east of Iceland, mid-June for MSW and early July for 1SW salmon.

2.2.2.6 Denmark

The Working Group collects and routinely reports the annual catch of salmon taken in Denmark. However, the small Danish catches are not included in the assessment process used in developing catch advice for the distant water fisheries.

Catch: The catch input is based on continuously collected reports of salmon taken in the recreational fishery in Danish west coast streams (from Internet sources), which all hold populations of wild salmon. In four of these, where salmon populations have always been found, there is a large proportion of reared (finclipped) salmon in the catch, but these are all F1 offspring from the original populations. In the one catchment in eastern Denmark (Gudenå), where the salmon population is not genetically native to the stream, the annual catch is guesstimated.

Level of unreported catch: Unreported catch is expected to be negligible in the western streams because the fishing is closely regulated and controlled by the anglers. In the eastern stream (Gudenå) unreported catch is guesstimated.

Exploitation rates: Exploitation rates may be derived from the total catch related to estimates of the total run (calculated by mark-recapture surveys on a three-year cycle in the four streams with original populations on the west coast).

2.2.3 Data inputs for Southern NEAC countries

2.2.3.1 France

Catch: The estimation of salmon catch in France comes from two main sources: (1) mandatory declaration of rod and line catches and from the Adour nets operating in the lower river (scales are sampled from each fish caught) to the Office National de l'Eau et des Milieux Aquatiques (ONEMA), under the Ministère de l'Ecologie, which assumed responsibility from the Conseil Supérieur de la Pêche (CSP) in 2006; and (2) mandatory declaration of catches made by professional net fishermen to Affaires Maritimes, under the Ministère de la Mer, who since 2008 have delegated responsibility for collection and first processing of catch data to the Regional Boards for Sea Fisheries and Aquaculture Catch. At the same time, catches at sea are declared to the Institut Français de Recherches pour l'Exploitation de la Mer (Ifremer), who are responsible for archiving and scientific processing of all fisheries data. Salmon and sea trout catches have not been reliably collated and made available until recently. Since 1985, the 1SW/MSW split has been based on scale interpretation of the in-river catch (based on scale reading) and on a categorisation based on length thresholds for catches in estuaries and at sea. The figures prior to 1985 are not considered as reliable as the later ones.

Level of unreported catch: Unreported legal catch for the rod and line fishery has been estimated by catch inquiries made by environmental inspectors of ONEMA on each river. These procedures are still operating in some areas, but estimates are considered less reliable in recent years. The estimation of the professional net fishery catch (Adour Basin) is thought to be reliable and no unreported legal catch is considered to apply.

For most years, the unreported illegal catch is not assessed and a minimal value is provided on a precautionary basis. This unreported illegal catch has been assessed in some years by ad hoc inquiries in the estuary of a number of rivers in Brittany (e.g. in 2001) and on the coast (e.g. Baie of Mont Saint-Michel in 2000). The “unreported catch” is included in the nominal catch. No estimates of unreported catch are available for the early part of the time-series (prior to 2001). Thus, the rates input to the model for 1SW and MSW for the early period are near zero and range from -0.00001 to 0.00001. Higher values in the range 20% to 40% for 1SW and 15% to 30% for MSW fish are applied more recently.

Exploitation rates: Exploitation rates are derived from the index River Scorff in Brittany. This is an in-river rate, by rods only, where there are no, or very few, fish thought to be caught on the estuary or coast. Rates are also derived for the Adour river system, where a rough estimation is provided by using the lower values of adult run estimates through facilities in the three rivers flowing to Adour, and the declared catches on the coast, estuary and river, respectively by nets and rods. Some caution is necessary regarding these rates from the Adour given the uncertainties in the different estimates. The rod catch on the index River Nivelle is very small and the probable net exploitation in the estuary and coast is unknown, so exploitation rates are not used for this system. Some data on exploitation rates are also collected by ONEMA on the index River Bresle, but sea trout are the dominant angled species in this river.

2.2.3.2 Ireland

Catch: The data are derived from annual declared catches within fisheries districts, management units implemented by Regional Fisheries Boards. Since 2007 river and estuarine specific angling and commercial catch data have been compiled. The Fisheries Boards were amalgamated into a single body, Inland Fisheries Ireland, in 2010 which currently takes responsibility for compiling catch statistics. Catches are split by age on the basis of a reported age distribution from 1980 to 1988. In the absence of any other information the mean proportion of 2SW salmon in the series (7.5%) has been used since 1988 and a mean of 10% has been used prior to 1980. Since the introduction of a carcass tagging and logbook scheme for angling and commercial fisheries in Ireland in 2002, sea age classes in the time-series since 2007 have been determined based upon catch dates and weights in accordance with national river stock assessments. The catch is not corrected for returns from releases of smolts for ranching or enhancement but these are not a major component of the catch.

Level of unreported catch: The values are guesstimated from local reports and knowledge achieved during catch sampling and fisheries protection activities.

Exploitation rates: A coded-wire tagging (CWT) programme has been operated in several rivers in Ireland since 1980. Up to 300 000 hatchery smolts and up to 5000 wild smolts are tagged and released annually. There is also a substantial dataset on wild salmon from the monitored River Burrishoole, providing a further index of wild returns and exploitation rates. Overall, there are estimates of exploitation rates available for three wild stocks and seven hatchery stocks for both 1SW and 2SW salmon. Up to the closure of the marine mixed-stock fishery in 2006, the annual mean of the 1SW wild exploitation index is used as the input data for the lower range of exploitation in the PFA model while the mean of the 1SW hatchery index is used as the upper range. The annual mean of the 2SW wild and hatchery exploitation index was used as the input data for the upper and lower range of exploitation in the PFA model depending on which is higher or lower in that year. Since 2006 the main exploitation

input has been from the rod catch which is estimated from coded wire tags estimates for some rivers and also rivers with counters.

2.2.3.3 UK (England & Wales)

Catch: Nominal catches for UK (England & Wales) have been derived from the catch returns submitted by netsmen and anglers and split into 1SW and MSW categories using two different methods. Since 1992, monthly age-weight keys derived from salmon caught in the River Dee trap (an index river) have been used to estimate the age of all rod-caught fish where a weight and date of capture have been provided. This has then been scaled up to the total catch (rods and nets combined) on a pro-rata basis. In earlier years (1971–1991), the age composition of the total catch has been estimated using the mean weight of the fish caught and the mean weight of 1SW and MSW salmon recovered in tagging programmes.

As the contribution of farmed and ranched salmon to the national UK (England & Wales) catch is negligible, the occurrence of such fish is ignored in the assessments of the status of national stocks. However, a large proportion of the fish taken in the northeast coast fishery are destined for Scottish rivers, and these are deducted from the returning stock estimate for UK (England & Wales) and added to the data for UK (Scotland) in the ICES assessment. This proportion is estimated to have declined from 95% of the northeast net catch in the early part of the time-series to 75% in the late 1990s and to around 65% since 2003. This reflects both the steady improvement in the status of the stocks in northeast England and the phase out of the English driftnet fishery since 2003.

Level of unreported catch: All licence holders are required to provide the Environment Agency with details of their catch of salmon and migratory trout and the number of days fished on each river or, for nets, each fishery at the end of the season. Catch returns are received from all net licence holders and from ~90% of full season anglers, and the latter account for the majority of fish caught in a catchment, typically 96–98%. The main correction for underreporting is therefore currently made in respect of perceived inaccuracies in the returns, although more substantial corrections have applied in the past.

There are few independent measures of underreporting in the rod fishery, but these indicate that the level is currently small. A value of 10% is applied for correction purposes based on the method of Small (1991). Historically, underreporting was a much more serious problem. As a result of changes in the licensing and associated catch return system covering UK (England & Wales) in the early 1990s, the percentage of underreporting in the rod catch was estimated to have decreased from ~ 50% to ~20%. Since the mid-1990s, awareness campaigns and enhanced catch reminder systems have further reduced underreporting to the levels currently estimated.

For the net fishery, a figure of 8% has been used in recent years to adjust for the level of underreporting, based on the outcome of surveillance operations. The level may have been substantially higher in the past in certain fisheries, possibly as much as 50%. However, following the successful introduction of logbooks and a carcass tagging scheme in 2009, there is now considered to be minimal underreporting in net fisheries. A figure of 2% has been assumed since 2009.

An earlier questionnaire survey of Environment Agency enforcement staff suggested illegal catches were around 12% of the declared net and rod catch. However, since the introduction of a carcass tagging scheme and a ban on the sale of rod caught fish in 2009, it has been substantially more difficult to dispose of illegally caught fish. Since

this time, illegal catches have been estimated to have been reduced to 6% of the declared catch.

Exploitation rates: Exploitation rates for a number of monitored fisheries in UK (England & Wales) are derived annually. National exploitation rates have then been estimated by deriving time-series of 'standard fishing units' employed in the salmon fisheries for the period 1971 to the present. For the period 1971 to 1997, these are calculated from the numbers of licences issued weighted by their relative catching power, which is estimated from historic cpue data; and for the period 1998 to the present, they are calculated from the numbers of days fished by different net categories weighted in the same way. The annual exploitation rates are then estimated by referencing the number of 'standard fishing units' employed over the two periods relative to average age-specific exploitation estimates derived for the 1997 and 1998 seasons.

Additional information: Further details on the derivation of estimates within UK (England & Wales) are available in the annual stock status reports (e.g. Cefas, Environment Agency and Natural Resources Wales, 2014), available at:

<http://www.cefas.defra.gov.uk/publications/files/SalmonReport2013-final.pdf>

2.2.3.4 UK (Northern Ireland)

Area split: Originally, a single assessment was carried out for UK (Northern Ireland). However, the data used were derived from two fishery management areas (Foyle and Fishery Conservancy Board (FCB) areas), which publish separate catch statistics and have differing fishing regulations. On the basis that stock status in the two areas may differ (Crozier *et al.*, 2003) the two areas were assessed separately from 2001.

Catch: As no commercial fishing has been conducted in the Loughs Agency area since 2010 and in the DCAL area since 2012 the Northern Ireland catch statistics currently (since 2014) rely solely on rod catches. Overall UK (Northern Ireland) rod catch estimates are available since the introduction of a carcass-tagging scheme in 2001. These catch statistics are used as an input in the model. Estimates of sea age composition of the catch for the time-series are based on 1SW/MSW data from adults returning to the River Bush (an index river).

Level of unreported catch: Estimates of unreported catch, as a result of illegal fishing, are based on intelligence reports from DCAL and Loughs Agency fishery officers. These are guesstimates only, with no verification possible. Annual adjustments in unreported catches have been used since tagging programmes started in the mid-1980s. Prior to that, a constant under reporting figure is used, as no annual data are available. The introduction of the carcass tagging scheme in 2001 has led to a reduction in unreported catches.

Exploitation rates: Estimates of exploitation rates were historically based on the River Bush microtagging programme. Exploitation from this monitored river (which is in the DCAL fishery area) was used as an input figure for all UK (Northern Ireland) fisheries (Foyle and DCAL areas). However, as currently no commercial fishery for salmon exists in the DCAL and Loughs Agency areas, exploitation rates are based on rod exploitation in the DCAL and Loughs Agency alone.

Possible improvements: A possible improvement would be to have better data available on sea age composition of all Northern Irish fish. Currently the River Bush estimate is applied to all Northern Irish data, but independent data from large river systems like the Bann and Foyle would result in a more reliable country wide estimate.

In addition, a higher return rate for the carcass tagging scheme would result in more reliable estimates of exploitation rates. Recently the carcass tagging return rate in UK (Northern Ireland) has varied between 14% and 55%.

2.2.3.5 UK (Scotland)

Area split: The country is divided into eleven regions for the purposes of collating and publishing salmon and sea trout fishery statistics (Marine Scotland Science, 2012). Within the PFA run-reconstruction model, UK (Scotland) is divided into two broad areas (east and west), the split being influenced by the contrasts in climate, river size and the timing and sea age of returning fish. The east grouping comprises the east, northeast, Moray-Firth, and north statistical regions, the remaining regions comprise the west grouping.

Catch: Catches were collated according to the area split defined above. Reported retained catches of wild salmon, taken by both net and rod fisheries, are provided separately for two age classes, one sea-winter and multi sea-winter fish. Catch sampling programmes have shown that there is a variable (by region, year, and fishery) proportion of 1SW salmon categorised as MSW salmon in the reported catches. Catch data used in the model are not currently corrected to account for such misreporting.

Level of unreported catch: The ranges used in the national model are based on previous guesstimates made by local managers in some eastern areas of the country (MAFF, 1991). The differences in the ranges used for the east and west groupings are based on a subjective view that unreported catches in the west area are likely to be greater than in the east area. It is thought that human population densities are lower in the west and therefore there is likely to be less surveillance over the reporting of salmon catches. Further, west coast rivers are generally smaller and more numerous than east coast rivers, leading to a greater number of locations where unreported catches may be taken. Ranges provided are a subjective estimate of uncertainty in these parameters.

Exploitation rates: Rates for the national model are guesstimates derived from estimated exploitation rates over a range of fisheries in the early part of the time-series (Shearer, 1992), and the time-series of exploitation rates derived from the River North Esk (Marine Scotland Science, 2010), together with reported net effort indices and the proportion of rod caught fish which were subsequently released (e.g. Marine Scotland Science, 2013).

2.2.3.6 Spain

The Working Group collects and routinely reports the annual catch of salmon taken in the recreational rod fisheries in Spain (mainly Asturias). However, the small Spanish catches are not included in the assessment process used in developing catch advice for the distant water fisheries.

2.2.4 Data inputs for Faroes and West Greenland fisheries

2.2.4.1 Faroes

Reported catch: Catches are derived from the landings of salmon caught in the commercial and research fisheries that operated in the Faroes EEZ and the northern Norwegian Sea. Catches for each season (i.e. November in year n to May in year $n+1$) are assigned to the second year (i.e. year $n+1$). These fish are classified into 1SW and

MSW age groups according to their age (or potential age) on January 1st during the fishery (i.e. a post-smolt caught in November is classified as 1SW).

Unreported catch: All fish less than 63 cm total length have been discarded in this fishery and so an unreporting rate of 10–15% (with an error of $\pm 5\%$) has been used for 1SW fish; there is thought to have been negligible non-reporting of MSW fish.

Catch composition: Estimates of the proportion of farmed fish in the catch for the period 1981 to 1995 have been derived from scale reading (ICES, 1996; Hansen *et al.*, 1997); prior to 1981 all fish are assumed to have been wild, and since 1997 a value of 0.8 has been used. The country of origin of the catch has also been estimated based on tagging studies undertaken in the early 1990s (Hansen *et al.*, 1999) and applied to the full time-series of catches.

2.2.4.2 West Greenland

Catch: The total nominal catch (i.e. tonnes round fresh weight) in the West Greenland fishery is reported and converted to numbers using a mean weight obtained from the sampling programme.

Unreported catch: Estimates of unreported catch were not provided for the period from 1993 to 1999; an annual estimate of non-reported catch, varying from 5 to 20 tonnes was provided by the Greenland representative. Since 2000 a nominal figure of 10 t per year has been provided.

Efforts have been made to provide further information on the level of unreported catch at West Greenland. Since 2002, some assessment of the unreported catch, primarily for commercial landings, has been provided by comparing the weight of salmon seen by samplers involved in the international sampling programme and the corresponding community-specific reported landings. However, since sampling only occurs during a portion of the fishing season, these are considered to be minimum estimates for unreported catch. In addition, there is currently no quantitative approach for estimating the unreported catch for the private fishery. A telephone survey of fishers was carried out following the 2014 season and provisional findings were provided to WGNAS in 2015. These are discussed further in Section 5.1 of the Working Group report (see above); such investigations may provide a basis for revising estimates of unreported catch in future.

Continent of origin: The catch at West Greenland was divided into NAC and NEAC components using scale characteristics until around 2000 and since that time genetic analysis has been used. For the period when scale characteristics were used, the input data to the model are the minimum and maximum estimates of the proportion of NAC fish (from which minimum and maximum proportions of NEAC fish are calculated). For the subsequent period, the inputs are the numbers of NAC and NEAC fish identified in the samples.

2.2.5 Improvements to NEAC input data

NEAC countries have made ongoing efforts to improve the input data used in assessments. Modifications to input variables are reported by WGNAS in the year in which they are first implemented.

Over recent years, efforts have been made to reduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes). However, the methods used to derive estimates of unreported catch vary markedly between countries. For example, some

countries include only illegally caught fish in the unreported catch, while other countries include estimates of unreported catch by legal gear as well as illegal catches in their estimates.

Descriptions of the national approaches used for evaluating unreported catches have been reported at various WGNAS meetings (e.g. ICES, 1996; 2000; 2002; 2010a). In addition, detailed reports describing national procedures for evaluating illegal and unreported catch, and efforts to minimise this, were submitted by parties to NASCO in 2007 in support of a special theme session on this issue. Full details are available at: [http://www.nasco.int/pdf/2007%20papers/CNL\(07\)26.pdf](http://www.nasco.int/pdf/2007%20papers/CNL(07)26.pdf).

Input data commonly rely on rod catches and the practice of catch and release has become increasingly important in recent years to reduce levels of exploitation on stocks. In the NEAC areas, catch and release estimates from the rod fisheries are not available from all countries and, when they are, corrections for catch and release mortality are commonly not applied. As the practice of catch and release is increasing, WGNAS have previously recommended (ICES, 2010a) that consideration should be given to incorporating mortality associated with this practice in river-specific, regional and national assessments.

The procedures currently used to incorporate catch and release and unreported catches into regional, national and international assessments are summarised in Appendix 1 (from ICES, 2010a).

One weakness of the NEAC model is that it is heavily dependent upon catch data and estimates of exploitation rate. In most salmon fisheries in the NEAC area, more than half the catch is reported, and in many cases it approaches 100%. However, as stocks have declined, exploitation rates have been reduced to very low levels, and estimates of abundance are therefore becoming increasingly sensitive to this parameter. This inevitably means that uncertainty in the estimates is increasing, and it therefore strengthens the need to make use of alternate sources of information on stock abundance, such as adult counts.

2.3 Input data for assessments–NAC area

The run-reconstruction model for NAC developed by Rago *et al.* (1993a) is used to estimate the PFA of non-maturing 1SW salmon of North American origin (beginning in 1971). Only the West Greenland fishery is of relevance in the context of distant water exploitation of NAC stocks. This fishery exploits predominantly (>95%) 1SW non-maturing salmon (destined to return primarily as 2SW salmon) and hence it is only necessary to estimate the abundance of this age group prior to the fishery at Greenland. The other fish taken in the fishery represent 2SW and older non-maturing salmon and previous spawners (ICES, 2003). However, PFA estimates for maturing 1SW salmon as well as large salmon (containing all MSW age groups of salmon including repeat spawners) can be derived from the run-reconstruction model.

The starting point for the reconstruction requires estimation of the returns of 2SW salmon to the six regions in eastern North America: Labrador, Newfoundland, Québec, Gulf, Scotia-Fundy, and USA. With the progressive closure of commercial fisheries (1984 for the Gulf and Scotia-Fundy regions; 1992 for Newfoundland; and 1998–2000 for Labrador and Québec) abundance estimates of 2SW salmon have relied less on harvests and increasingly on estimated returns to rivers raised to production areas. The returns for each region are estimated with the uncertainty defined by a range of minimum and maximum values based on the best information available for each region (Chaput *et al.*, 2005).

The annual pre-fishery abundance of non-maturing 1SW fish for year i , destined to be 2SW returns (excluding 3SW and previous spawners), represents the estimated number of salmon at West Greenland prior to the start of the fishery on August 1st. Definitions of the input variables used in the model are given in Table 2.3.1. The PFA estimate is constructed by summing 2SW returns in year $i+1$ [NR2($i+1$)], 2SW salmon catches in commercial and Aboriginal peoples' food fisheries in Canada [NC2($i+1$)], and catches in year i from fisheries on non-maturing 1SW salmon in Canada [NC1(i)] and Greenland [NG1(i)].

Table 2.3.1. Definitions of key variables used in continental run-reconstruction models for North American salmon.

| | |
|------------|-------------------------------------------------------------------------------------------------------------------------------|
| i | Index for PFA year corresponding to the year of the fishery on 1SW salmon in Greenland and Canada |
| M | Natural mortality rate (0.03 per month) |
| t_1 | Time between the midpoint of the Canadian fishery and return to river = 1 month |
| S_1 | Survival of 1SW salmon between the homewater fishery and return to river $\{exp^{-M \cdot t_1}\}$ |
| $H_{s(i)}$ | Number of "Small" salmon caught in Canada in year i ; fish < 2.7 kg |
| $H_{l(i)}$ | Number of "Large" salmon caught in Canada in year i ; fish ≥ 2.7 kg |
| AH_s | Aboriginal and resident food harvests of small salmon in northern Labrador |
| AH_l | Aboriginal and resident food harvest of large salmon in northern Labrador |
| f_{imm} | Fraction of 1SW salmon that are immature, i.e. non-maturing; range = 0.1 to 0.2 |
| af_{imm} | Fraction of 1SW salmon that are immature in native and resident food fisheries in N Lab |
| q | Fraction of 1SW salmon present in the large size market category; range = 0.1 to 0.3 |
| $MC1(i)$ | Harvest of maturing 1SW salmon in Newfoundland and Labrador in year i |
| $i+1$ | Year of fishery on 2SW salmon in Canada |
| $MR1(i)$ | Return estimates of maturing 1SW salmon in Atlantic Canada in year i |
| $NN1(i)$ | Pre-fishery abundance (PFA) of non-maturing 1SW + maturing 2SW salmon in year i |
| $NR(i)$ | Return estimates of non-maturing + maturing 2SW salmon in year i |
| $NR2(i+1)$ | Return estimates of maturing 2SW salmon in Canada |
| $NC1(i)$ | Harvest of non-maturing 1SW salmon in Nfld + Labrador in year i |
| $NC2(i+1)$ | Harvest of maturing 2SW salmon in Canada |
| $NG(i)$ | Catch of 1SW North American origin salmon at Greenland |
| T_2 | Time between the start of the fishery at West Greenland (August 1) and return to the coast of North America = 10 months |
| S_2 | Survival of 2SW salmon between August 1 (at West Greenland) and return to the coast of North America $\{exp^{-M \cdot T_2}\}$ |
| $MN1(i)$ | Pre-fishery abundance of maturing 1SW salmon in year i |

2.3.1 Data inputs for NAC

The latest data input variables used in running the NAC assessment are listed at Appendix 4. More detailed descriptions of how the model input data have been derived for each region of North America are presented below.

2.3.1.1 Labrador

For Labrador stocks, it was thought inappropriate to develop total recruits from angling catches and exploitation rates similar to techniques used for rivers in insular Newfoundland. The problem with using angling catches to derive returns for Labrador is, that until 1994, there were no estimates of exploitation rates available other than for the salmon population of Sand Hill River and these were 20 years out of date. Also, because Labrador coastal rivers are isolated, the exploitation rates are low and highly variable depending on the presence of an angling camp and its success in attracting guests as well as the nearness of local communities. Thus, exploitation rates would depend and vary from one year to the next based on the success of angling camps in attracting anglers and may not be applicable to other Labrador rivers. Thus, all estimates of returns and spawners until 1998 were based on commercial catches as the only source available of usable continuous time-series of data.

Before 1998

The general approach is to use exploitation rates to convert commercial catches of small and large salmon in Labrador to total population prior to the commercial fishery. River returns and spawners were estimated by subtracting the commercial catch from these populations, and accounting for non-Labrador interceptions. The estimated number of Labrador origin large returns is calculated as:

$$LR = (CC \cdot PL) / u \quad (1)$$

where,

LR = Labrador returns, PL = proportion Labrador origin, CC = commercial catch, and u = exploitation rate

The estimated number of Labrador origin small returns is determined from equation (1) but using commercial catches of small salmon.

Parameter values for sea age and the proportion of salmon of Labrador origin comes from the sampling program in the commercial fishery, 1974–1991. In 1997, commercial sampling resumed with samples being collected throughout the fishery at Makkovik and Rigolet in SFA 1 and Cartwright and St Lewis/Fox Harbour in SFA 2. River age distribution of commercial samples of small and large salmon from Labrador have been found to consist, on average, of about 75–80% river age 4 and older in SFAs 1 & 2. The commercial samples came from commercial catches sampled in Labrador at several sites along the Labrador coast including Square Islands (SFA 2) and at Nain (SFA 1) (Anon, 1993b). In total, 46 320 salmon were sampled for scales and aged. Labrador salmon stocks are thought to contribute about 70% of the total production of four year, and older, river age salmon, with the other 30% coming from northern Québec. Thus, when non-Labrador salmon are factored in at 30% applied to the river age distribution, then 60–80% of the harvest of small and large salmon (PL) in Labrador are of Labrador origin (Anon, 1993b). In 1997, in SFA 1, the percentage of the commercial catch that was of Labrador origin was for large salmon 68% (95% C. I. 64.3–72.5%); whereas for small salmon it was 39% (95% C.I. 35.6%–41.6%). In 1997, in SFA 2, the percentage of the commercial catch that was Labrador origin was for large salmon 92% (95% C. I. 88.4–95.2%); whereas for small salmon it was 80% (95% C.I. 74.8%–85.0%).

Exploitation rates (u) were calculated from the smolt tagging study in 1969–1973 on Sand Hill River (Reddin, 1981; Reddin and Dempson, 1989). Exploitation rates of 0.28 to 0.51 for small salmon and 0.83 to 0.97 for large salmon from the tagging study were

changed to base exploitation rates of 0.3 to 0.5 on small salmon and 0.7 to 0.9 on large salmon and were assumed to apply to all of the salmon populations in SFAs 1, 2, and 14B for the period of 1969–1991 (Anon., 1993b). After 1991, due to the Management Plans for the commercial fishery in Labrador and Newfoundland, several changes occurred that would reduce exploitation of Labrador origin salmon. These changes include: (1) reductions in effort as commercial salmon fishermen chose to sell their licences from a buy-out agreement begun in 1992, (2) a moratorium on commercial fishing in Newfoundland would increase the number of Labrador salmon in Labrador coastal waters, and (3) season reductions due to the varying opening dates and early closures from the quotas applied in 1995 and 1996. The effects of these changes were quantified in the exploitation model as follows:

$$u=1-e^{-aF} \quad (2)$$

where: a = fraction of the 1991 licensed effort remaining in 1992–1996.

In 1994–1996, the licensed effort for all of Labrador was 37% of the 1991 level of 570 licences, in 1993 it was 55%, and in 1992 it was 87%. In any given year, it was assumed that 90% of licensed fishermen were active. Fishermen reported during public consultations that in 1995 and 1996 many licensed salmon fishermen did not fish for salmon in 1995–1996 but fished for crab instead. This was verified by Fisheries Officers who reported that of the 218 licensed salmon fishermen only 132 were active in 1996. Another method of obtaining actual effort information is also available since, beginning in 1993 commercial fishing vessel (CFV) numbers have been recorded on sales receipts issued to fishermen by fish plants. Enumeration of licensed salmon fishermen actively fishing was made by determining the number of CFVs in the Statistics Branch catch records. Active effort in 1991 and 1992 was assumed to be 90% as it was in 1993 and 1994 from the CFV file. Thus, the exploitation rates (u) were modified due to effort reductions in equation (2) using estimated active licences from 1991 as a base and the number of active licences in 1995, 1996 and 1997. The modified exploitation rates (ue) for 1992–1997 used the licensed effort in equation (2).

The tagging study on Sand Hill River, 1969–1973 showed that Labrador small and large salmon were not only caught in Labrador, but also in the commercial fisheries along the northeast coast of Newfoundland (both small and large) and at West Greenland (large only) (Anderson, 1985). For small salmon, out of a total of 100 (1SW) tag returns there were 24 from Newfoundland. For large salmon, out of a total of 137 (2SW) tag returns there were 41 from Newfoundland.

For 1992–1997: the moratorium on commercial fishing in Newfoundland would have released small and large salmon to Labrador. The effect of salmon released from Newfoundland in 1992–1996 was evaluated against the exploitation rates as follows:

$$\begin{aligned} u_n &= (1 - ((24 * (1 - ue))/100)) * ue, \text{ for small salmon, and} \\ u_n &= (1 - ((41 * (1 - ue))/137)) * ue, \text{ for large salmon} \end{aligned} \quad (3)$$

The new estimates of fishing mortality (u_n) in 1992–1994 included adjustments for the closure of the commercial fishery in Newfoundland based on the results of the Sand Hill River tagging study. Season reductions due to the varying opening dates and early closures from the quotas applied in 1995 and 1996. In 1995, adjustments were made to account for the new opening date for the commercial fishery in Labrador of July 3 changed from June 20 the previous year. For 1995, the accumulative effect of these, weighted to SFA catches, was to reduce the catch so that for small salmon the current catch represents 86.0% of small salmon and 62.7% of large salmon. In 1996, the opening

date reverted to June 20 but the quota levels resulted in early closures in SFA 2 of 2A - July 10, 2B - July 8, and 2C - July 2 while SFA 1 and 14B did not close. For 1996, the accumulative effect of these weighted to SFA catches was to reduce the catch so that for small salmon the current catch represents 53% of small salmon and 61% of large salmon. In 1997, the opening date remained at June 20 but the quota levels resulted in early closures in SFA 2 of 2A - July 12, 2B - July 15, and 2C - July 13 while SFA 1 closed on October 15 as the quota was not caught. For 1997, the accumulative effect of these early closures was to reduce the catch so that for small salmon the current catch represents 47% of small salmon and 64% of large salmon. The season changes reduce catches and hence lower exploitation rates. The effect of shorter seasons in 1995, 1996 and 1997 was evaluated against the exploitation rates in section B as follows:

$$\begin{aligned} US &= UN * SC, \text{ for small salmon, where SC is season change, and} \\ US &= UN * SC, \text{ for large salmon} \end{aligned} \quad (4)$$

The new estimates of fishing mortality including effort reductions, adjustments for the closure of the commercial fishery in Newfoundland, and shorter seasons due to opening dates and quotas results in the following exploitation rates which were applied to catches. The cumulative effect of factors A, B, and C is to reduce exploitation on Labrador origin salmon.

Labrador origin 2SW returns (LR2SW) were derived from eq. 1 by:

$$LR2SW = LR * P2SW \quad (5)$$

where: P2SW = proportion of the large salmon that is 2SW salmon.

The SR1SW were calculated as in equation (5) but using P1SW which is the proportion of the catch that is 1-sea winter in age and maturing to enter freshwater and spawning in the year of capture. The parameter values for P1SW of 0.1 to 0.2 come from Anon. (1991).

The 2SW component was estimated separately for salmon caught in SFA 1, 2 and 14B. In SFA 1, commercial sampling at Nain of large salmon showed the proportion of 2SW were on average about 84% (n=6542), 1977–1991. Thus, a range of 0.7–0.9 was used for SFA 1. In SFA 2, commercial sampling of large salmon averaged 69% (n=4793) 2SW salmon, 1977–1991. There were no commercial samples available for SFA 14B. Thus, for SFAs 2 & 14B a range of 0.6–0.8 was used. For the 1SW component, commercial samples at Nain in SFA 1 of small salmon showed the proportion of 1SW salmon were on average about 94% (n=4757). In SFA 2 the 1SW component was on average about 97% (n=8872) of small salmon. There were no samples from commercial sampling in SFA 14B. In 1997, aged commercial samples indicated that the previous range was acceptable.

Total river returns of 2SW salmon (TRR) were calculated as follows:

$$TRR = LR2SW / (1-us) \quad (6)$$

The total river returns of small salmon are also calculated by equation 6 but from SR.

Spawning escapement (SE) or spawners was calculated according to the formula:

$$SE = TRR - AC, \quad (7)$$

where:

AC = angling catch which includes retained catch plus 10% of hook & released mortality for released salmon.

A couple of modifications were made to the estimation procedure for Labrador in 1997. Firstly, determination of exploitation rates were calculated separately for SFA 1, 2 and 14B using the active effort individually for each SFA. For SFA 1, the active number of licences declined from 141 in 1991 to 39 in 1997. For SFA 2, the active number of licences declined from 320 in 1991 to 99 in 1997. For SFA 14B, active licences declined from 52 in 1991 to 0 in 1997 when the fishery was closed. Exploitation rates determined as in equations 2, 3 and 4 are: SFA 1 - small was 0.0735 to 0.1399 and - large was 0.2221 to 0.3959; and SFA 2 - small was 0.0384 to 0.0728 and - large was 0.1589 to 0.2799.

Numbers of small and large salmon for SFAs 1 & 2 were estimated from the exploitation model while for SFA 14B the results of assessments on Forteau Brook and Pinware River were expanded to include all of the watershed in SFA 14B. Returns to SFA 14B were 663 to 1545 small salmon and 146 to 327 large salmon.

Total mortalities of small and large salmon were accounted for by summing commercial catches of small salmon in Labrador and Newfoundland, large salmon in Labrador, Newfoundland, and Greenland, angling catches in Labrador of small and large salmon including 10% of the hook and released salmon, and small and large spawners. All of the above mortality estimates except catches of Labrador salmon in Newfoundland, 1969–1991 and Greenland could be obtained from equations 1 to 7. Catches in Newfoundland and Greenland were assessed as follows:

Greenland: for 1969–1992 and 1995–2004, removals of Labrador salmon by the Greenland fishery were assessed from data based on the sampling program in commercial fish plants at West Greenland (Anon, 1996). The Greenland fishery catches salmon that would have returned to homewaters as large salmon in the year following the Greenland fishery. Numbers of Labrador salmon were determined by converting catches in kg to numbers of salmon of 1SW North American origin that were of river age 4 and older. The number of Labrador salmon were estimated by assuming that 70% of the production of 4-year and older river age salmon are from Labrador (Anon, 1993b).

Newfoundland: for 1969–1991, catches of Labrador small and large salmon in Newfoundland were included in total mortalities as the product of the ratio of tags caught in Newfoundland to Labrador and the catch in Labrador. For small salmon the ratio was $(24/(100-24)) = 0.32$ and for large salmon it was $(41/(137-41)) = 0.43$.

1998–2001

For the years, 1998–2001 when only one or two counting projects took place in Labrador, the raising factors previously used and explained of 1.04 to 1.49 for small salmon and 1.05 to 1.27 for large salmon were used to estimate returns and spawners for Labrador from the overall PFA minus catches in Greenland, as was the case in previous years. However, in this case returns to rivers were derived for Labrador by subtracting landings in food fisheries. Also, catches in 1994–2006 were updated to reflect changes made to catch statistics in Labrador from the Licence Stub Return System. Procedures for the collection and compilation of commercial and angling fishery data are described in Ash and O'Connell (1987) for fishery years 1974–1996. For years 1969–1974, commercial catch data came from Anon. (1978). In 1997, the angling catch statistics were converted to a licence stub system (O'Connell *et al.*, 1998) which continues to present.

2002-present

Counting projects occur on four Labrador rivers; out of about 100 extant salmon rivers. Because they occur on the same four rivers each year, it is possible to extrapolate from return rates for small and large salmon per accessible drainage areas in these four rivers to unsurveyed ones in the remainder of Labrador. The area accessible drainages were 9267 km² for Lake Melville (SFA 1A), 25 485 km² for Northern Labrador (SFA 1B), 28 160 km² for Southern Labrador (SFA 2), and 2651 km² for the Straits Area (SFA 14B). Accessible drainage area in the counting facility rivers was 1878 km² resulting in an expansion factor of 35 to one. Not all rivers in Lake Melville were included due to a lack of information on presence of salmon populations in rivers in this region of Labrador. Lake Melville rivers whose drainage areas were included are Sebaskachu, Cape Caribou, Goose, MacKenzie, Kenamu, Caroline and Traversspine.

Return rates for SFAs 1A and 1B were derived from English River return rates with maximum and minimum values developed using the observed variability of return rates in SFA 2. Total returns and spawners for Labrador are estimated by Monte Carlo simulation based on 10 000 random draws from the range of values assuming return rates per km² of accessible drainage were uniformly distributed. The return rates for each SFA were then multiplied times the total accessible drainage area to derive total returns of small and large salmon. Ranges of values were developed to convert numbers of small and large salmon to numbers of 1SW and 2SW salmon from scale age information collected from counting fences and angling fisheries in Labrador. A bootstrap procedure was used to develop estimates of the proportions of sea age 1 salmon in estimates of small salmon returns and spawners, proportions of sea age 2 salmon in estimates of large salmon returns and spawners and proportions of sea age 1 salmon in the estimates of large salmon returns.

Sea age correction factors were:

- Small to 1SW - 96 to 100%
- Large to 2SW - 60 to 71%
- Small overlap in large - 12 to 21%

Spawners of 1SW and 2SW salmon were derived by subtraction of angling catches including an estimate of hook and release mortalities (10%) from the returns.

2.3.1.2 Newfoundland

Inputs for the run-reconstruction model for Newfoundland include estimates of small, large and 2SW returns and spawners to rivers (minimum and maximum). The methods used to estimate returns and spawners to the rivers in Newfoundland are described by Reddin and Veinott (2010). In brief, returns and spawner estimates were derived from recreational fishery exploitation rates of retained small salmon for rivers with enumeration facilities; and ratios of large to small salmon were utilized to estimate large salmon. Exploitation rates were then applied to all rivers with reported angling catches. A non-parametric bootstrap technique was used, whereby exploitation rates and ratios of large to small salmon from rivers with enumeration facilities were chosen at random with replacement. The 95th confidence interval from 500 iterations of the weighted exploitation rate and ratio of large to small salmon was applied to angling catches on a Salmon Fishing Area (SFA) basis. The midpoint of the 95th confidence interval was used as the minimum and maximum estimate returns of large and small salmon in each SFA. Estimates of 2SW returns are based on the expected proportion of 2SW in the large salmon category (≥ 63 cm). Commercial and recreational angling

catches were derived as described for Labrador (2.3.1.1). Spawners in all years were determined as the returns to rivers minus angling catches including an adjustment for hook-and-release mortality.

2.3.1.3 Québec

In order to estimate abundance of stocks, rivers were classified into six categories (C1–C6) depending on the information available to estimate salmon returns (according to the method of Caron and Fontaine, 1999), with C1 being the most reliable evaluation and C6 the least. C1 corresponds to a river where the evaluation of the returns is based on a counting method, either from a fence or from a visual count through snorkelling or from a canoe. C2 uses the same evaluation, but without knowing the number of small and large salmon, which is then estimated from proportions reported in the sport fishing landings and, if necessary, the catch and release. Salmon returns on C3 rivers are determined based on multiple correlation factors, using catch number, fishing effort, season duration and river accessibility distance (Guillouët, 1993).

When estimation of the returns using a C1–C3 category is not possible, and when data of returns from previous years are available, the C4 category is used. C4 assumes that interannual variations in salmon returns in the targeted river are approximately the same as variations observed in the other rivers of the corresponding region. Category C5 is for rivers where only landings data are available. In these rivers the salmon run is estimated from the average regional exploitation rate. Finally, a few small rivers have essentially no available data. C6 then assumes that the run is related to the available river salmon habitat and is estimated with respect to rivers of the same area for which run estimates and salmon habitat area are known. Estimated numbers of returns from C4 to C6 cannot be used as management tools regarding the conservation limit. However, they provide at least approximate numbers to estimate returns and spawners for salmon rivers in Québec.

The evaluation of the uncertainty associated with return estimates depends on the river category. For C1 and C2 rivers, the correction factor for the minimum and maximum number of returns is +5% and +10% for all rivers with a fish ladder and for all others in zones Q1 to Q3 and Q10. The correction factor for rivers with darker water from zones Q5, Q6 and Q7 is rather +10% and +30%. For the other categories, an uncertainty of $\pm 25\%$ is associated with salmon return estimates, except for category C3 where calculation depends on the method of Guillouët (1993).

The number of spawners is obtained using the return estimate minus all river catches, which include landings and other types of removal. In most cases, river catches include landings from sport fishing only, which may be conducted by native people such as that on the Betsiamites River. The other types of removal are of limited number and include mainly natural mortality, salmon captured for hatchery use and subsistence fishing when practised in river.

Overall return estimates for all Québec rivers are obtained by adding in-river salmon returns, commercial fishing (when operated), native people subsistence fishing when practised in estuaries and an estimate of non-registered landings. However, little scientific data are available on non-registered landings and thus, estimates are based on good judgment, following consultations with regional biologists.

2.3.1.4 Gulf

Estimation of returns and spawners are developed for the four salmon fishing areas of Gulf Region (SFAs 15 to 18).

SFA 15

The major river in this area is the Restigouche River. The returns and spawners are estimated for the Restigouche River exclusive of returns to the Matapedia River, which are included in Québec zone Q1. The Restigouche River stock assessment is based on angling catch with assumed exploitation rates between 30% (min.) and 50% (max) with estuary catches added back after the estimates of returns. Return and spawner estimates for SFA 15 are based on Restigouche River data, scaled up for SFA 15 using angling data. The return and spawner estimates for SFA 15 are derived from the return and spawner estimates for Restigouche (New Brunswick). The minimum and maximum return and spawner estimates are derived from the minimum and maximum ratios of angling catch in all of SFA15 relative to angling catch in Restigouche (New Brunswick) (min = 1.117; max = 1.465). Harvests represent retained angling catch plus 6% catch and release mortality for released fish. The proportion of 2SW in large salmon numbers is based on aged scale samples from angling, trapnets, and broodstock. In the years when no scale samples analysis is available, a mean value of 0.65 is used.

SFA 16

The most important Atlantic salmon river in SFA 16 is the Miramichi River. The Miramichi makes up 91% of total rearing area of SFA 16 and returns to the river are assessed annually. For 1971 to 1991, minimum and maximum values are based on capture efficiencies of the Millbank estuary trapnet representing a lower CI of -20% of the estimate and upper CI of 33% of the estimate. For 1992 to 1997, minimum and maximum are lower and upper CI and based on estimate bounds of -18.5% to +18.5%. Since 1998 to the present, minimum and maximum are 5th and 95th percentile range from a Bayesian hierarchical model used in the assessment. Returns to SFA 16 are Miramichi returns (Minimum, Maximum) / 0.91. Proportion 1SW in small salmon is from scale ageing; proportions have varied from 0.97 to 1.0. Proportion 2SW in the large salmon category is obtained from scale ageing. Spawners are returns minus harvests. For 1998 to 2011, the harvest of large salmon is estimated as the sum of the aboriginal fisheries harvests for large salmon and 1% of the large salmon catch (30% exploitation rate, 3% catch and release mortality). The harvest of small salmon is estimated as 30% of the small salmon return plus the harvest from the aboriginal fisheries.

SFA 17

For 1970–1994, small returns are estimated from retained small salmon catch in the Morell River divided by the river-specific exploitation rate. Salmon catch in the Morell River was estimated in 1970–1990 by DFO Fisheries Officers; and in 1991, 1992, and 1994 by angler mail-out surveys. The number of small retained salmon in 1993 was not recorded, so the number used is the mean for 1986–1992. For 1970–1993, exploitation rate was taken as the mean of exploitation rates estimated for 1994, 1995, and 1996 (0.317). For 1994, exploitation rate was 0.34. The min and max of small returns are calculated using exploitation ± 0.1 ; e.g. 0.34 ± 0.1 gives 0.24 and 0.44. Large returns = (number of small returns/proportion small) - number of small returns. For 1970–1980, proportion small is calculated from numbers of small and large salmon in the angling catch of each year. For 1981–1994, proportion small is taken from counts at the Leards Pond trap on the Morell River. Small spawners = number of small recruits - number of small retained. Large spawners = number of large recruits - number of large retained.

Spawner estimates for 1995 to the present are derived from redd counts in 23 rivers. For years and rivers in which redd counts are unavailable, redd numbers are estimated by linear interpolation from the preceding and succeeding count year. Redd numbers in years prior to the first count are taken as the first count. Redd numbers in years after the last count are taken as the last count. Female spawners are estimated from the ratio of 3.357 redds/female spawner, measured in the West River in 1990. Total spawners are estimated from size-specific sex ratios derived from counts at Leards and Mooneys Ponds, Morell River, in 1986–2001. The proportion of salmon that are large is taken as 0.5 in the Cains, Carruthers, Trout (Coleman), Morell, Cardigan, West, and Dunk Rivers, and 0.9 in all other rivers. Spawners are presented as Min (estimated spawners -20%) and Max (estimated spawners + 20%). Returns are spawners + total estimated fishing mortality, including angler catches, hook-and-release mortality, and native harvests. Angler catches and hook-and-release mortality are estimated from angler card surveys. Returns are presented as Min (estimated returns -20%) and Max (estimated returns + 20%). It is assumed that large salmon and 2SW salmon are equivalent.

SFA 18

Returns and spawners to SFA 18 are derived from estimates of returns and spawners to the Margaree River, adjusted for the ratio of the SFA 18 angling catch to the Margaree River catch. For small salmon, the ratio of SFA 18 catch to Margaree catch varies between 1.15 and 2.71 for years 1984 to 2004. For large salmon, the ratio of SFA 18 catch to Margaree catch varies between 1.08 and 2.32 for years 1984 to 2004. Returns to Margaree River are estimated using various techniques.

- 1970 to 1983 angling catch divided by range of exploitation rates with maximum exploitation rate of 0.37 and minimum exploitation rate of 0.215;
- 1984 to 1986 based on annual assessments;
- 1987 to present angling catch and effort data from logbooks and provincial licence stubs are used to derive the returns. The catchability coefficient per rod day is estimated from angling catch and effort data for the years 1988 to 1996 when mark and recapture programmes were used to estimate returns, independently from angling data.

Spawners for 1970–1983 equal returns minus removals. Spawners for 1984 to the present equal returns minus catch for small salmon and returns minus catch, corrected for 5% mortality, for large salmon. 2SW salmon represent between 0.77 and 0.87 of large salmon returns and spawners.

2.3.1.5 Scotia-Fundy

Salmon originating in rivers of the Atlantic coast of Nova Scotia and southwest New Brunswick in Salmon Fishing Areas (SFAs) 19–21 and the portion of SFA 23 outside the inner Bay of Fundy comprise the Scotia-Fundy stocks. With the exception of at least one stock in SFA 19, they have a large salmon component that migrates to the North Atlantic/Labrador Sea (Amiro *et al.*, 2008). Estimates of returns and spawning escapement for the Scotia-Fundy stocks are provided as inputs to the run-reconstruction model. Inner Bay of Fundy Atlantic Salmon (SFA 22 and part of SFA 23) have been federally listed as endangered under the Canadian Species at Risk Act and are not included as inputs into the run-reconstruction model. With the exception of one population, inner Bay of Fundy stocks have a localized migration strategy while at sea and an incidence of maturity after one winter at sea.

Consistent with the requirements of the model, a range (minimum to maximum) of returns and spawning escapement for the Scotia-Fundy stocks is provided for the run-reconstruction model. The methods used to estimate total returns and spawners are described by Amiro *et al.* (2008). In brief, for SFAs 19–21, the escapement is based on the count of small and large salmon at the Morgan Falls fish-way on the LaHave River from 1970 to the present year, scaled up to the region using the relationship between this count and the recreational catch data for rivers in SFA 19 to 21 from 1970 to 1997 and a catch rate for the LaHave River from 1970 to 1997. Estimates of the returns also include estimates of landings in the commercial salmon fisheries in SFA 19–21 from 1970 to 1983. The model is fitted using maximum likelihood, and the 90% confidence limits are carried forward as the minimum and maximum values. In SFA 23 from 1970 until 1992, estimates of total 1SW and large wild-origin salmon returns are based on the estimated number of returns destined for tributaries above Mactaquac Dam on the Saint John River; this includes in-river and outer-Fundy commercial landings (1970–1971 and 1981–1983), in-river aboriginal harvests (since 1974), and counts at Mactaquac Dam. These estimates are raised by the proportion of the total accessible productive habitat in SFA 23 that is upstream of Mactaquac Dam (0.4–0.6). Hatchery-origin returns were attributed to above Mactaquac Dam only and no hatchery 1SW and MSW returns were estimated for other rivers within SFA 23 (outer Fundy). Since 1993 the estimates of 1SW and MSW returns to the Nashwaak River have been used to estimate the wild production from tributaries of the Saint John River below Mactaquac Dam. The estimated 1SW and MSW returns to the Nashwaak River (above Counting Fence), is raised by the proportion of the total production area accounted for below Mactaquac (0.21–0.3) and then added to the above Mactaquac totals.

2.3.1.6 USA

Total salmon returns and spawners for USA rivers are based on trap and weir catches and for the small rivers in Maine that do not have fish counting facilities, estimates of spawners were based on redd counts.

2.3.2 Improvements to NAC input data

Modifications to input variables used in assessments for the NAC area are reported by WGNAS in the year in which they are first implemented.

Over recent years, efforts have been made to reduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes). However, the methods used to derive estimates of unreported catch vary markedly between countries. For example, some countries include only illegally caught fish in the unreported catch, while other countries include estimates of unreported catch by legal gear as well as illegal catches in their estimates.

Descriptions of the national approaches used for evaluating unreported catches have been reported at various WGNAS meetings (e.g. ICES, 1996; 2000; 2002; 2010a). In addition, detailed reports describing national procedures for evaluating illegal and unreported catch, and efforts to minimise this, were submitted by parties to NASCO in 2007 in support of a special theme session on this issue. Full details are available at: [http://www.nasco.int/pdf/2007%20papers/CNL\(07\)26.pdf](http://www.nasco.int/pdf/2007%20papers/CNL(07)26.pdf)

Input data commonly rely on rod catches and the practice of catch-and-release has become increasingly important in recent years to reduce levels of exploitation on stocks. As the practice is increasing, WGNAS have previously recommended (ICES,

2010a) that consideration should be given to incorporating mortality associated with this practice in river-specific, regional and national assessments.

The procedures currently used to incorporate catch-and-release and unreported catches into regional, national and international assessments are summarised at Appendix 1 (from ICES, 2010a).

2.4 Biological and other data requirements

As noted previously, many of the 'conventional' data requirements (e.g. marine survey data and commercial cpue) used in the assessment of other commercially important fish species are inappropriate to salmon. A range of biological, catch and exploitation rates and other data pertinent to appropriate stock assessments are however, collected and made available to WGNAS to help inform assessments and to aid in responding to the various questions posed by NASCO.

Appendix 2 of this Stock Annex provides an overview of current and possible future data requirements for Atlantic salmon assessment/ scientific advice. This was compiled at a recent meeting of WGNAS (ICES, 2013) in relation to monitoring requirements under the European Data Collection Framework (DCF) and following a more detailed review of the data requirements under DCF (ICES, 2012c). This table illustrates the type of information collected/available, but is provided for illustrative purposes only. It should be noted that many Atlantic salmon producing countries fall outside the DCF provisions, which only relate to countries within the European Union. Further, Sovereign states are responsible for the regulation of salmon fisheries within their areas of jurisdiction. Formal ICES catch advice is only required for the distant water salmon fisheries, which take salmon originating in rivers of another party.

3 Assessment methods

In managing Atlantic salmon fisheries, NASCO has adopted a fixed escapement strategy (Potter, 2001), in recognition of the importance of the spawning stock to subsequent recruitment. Therefore, in managing the distant water fisheries at Faroes and West Greenland, the spawning requirements of the rivers contributing to these fisheries must be defined. Management advice, expressed as allowable harvest (tonnes), is then predicated on a forecast of salmon abundance prior to the fishery such that the spawning requirements of the contributing stocks can be achieved. The provision of catch advice thus proceeds through a number of steps:

- The definition of spawning objectives;
- The development of a measure of abundance prior to the fishery; i.e. the pre-fishery abundance or PFA;
- A measure of the spawning stock contributing to the PFA;
- A model to forecast the PFA;
- The development of a risk analysis framework for the catch advice.

These steps are described in detail in the following sections, subdivided as necessary for the different distant water fisheries and the various stock complexes which contribute to the two fisheries (Greenland and Faroes).

3.1 Definition of spawning objectives

3.1.1 Management objectives and reference points

Conservation limits (CLs) for North Atlantic salmon have been defined by ICES as the stock level that will achieve long-term average maximum sustainable yield (MSY). NASCO has adopted the following definition of CLs (NASCO, 1998): 'The CL is a limit reference point; having populations fall below these limits should be avoided with high probability.'

Atlantic salmon has characteristics of short-lived fish stocks; mature abundance is sensitive to annual recruitment because there are only a few age groups in the adult spawning stock. Incoming recruitment is often the main component of the fishable stock. For such fish stocks, the ICES MSY approach is aimed at achieving a target escapement ($MSY B_{escapement}$, the amount of biomass left to spawn). No catch should be allowed unless there is a high probability that this escapement can be achieved. The escapement level should be set so there is a low risk of future recruitment being impaired, similar to the basis for estimating B_{pa} in the precautionary approach. In short-lived stocks, where most of the annual surplus production is from recruitment (not growth), $MSY B_{escapement}$ and B_{pa} might be expected to be similar.

It should be noted that this is equivalent to the ICES precautionary target reference points (S_{pa}). Therefore, stocks are regarded by ICES as being at full reproductive capacity only if they are above the precautionary target reference point. This approach parallels the use of precautionary reference points used for the provision of catch advice for other fish stocks in the ICES area.

For the assessment of the status of stocks and advice on management of national components and geographical groupings of the stock complexes, where there are no specific management objectives:

- ICES requires that the lower bound of the 90% confidence interval of the current estimate of spawners is above the CL for the stock to be considered at full reproductive capacity (equivalent to a probability of at least 95% of meeting the CL).
- When the lower bound of the confidence limit is below the CL, but the midpoint is above, then ICES considers the stock to be at risk of suffering reduced reproductive capacity.
- Finally, when the midpoint is below the CL, ICES considers the stock to be suffering reduced reproductive capacity.

Ideally, Atlantic salmon should be assessed and managed on the basis of river-specific stock units, the scale corresponding best to the spawner to recruitment dynamic (Chaput, 2012). In reality, this is not the case for the majority of rivers, although efforts are continuing to develop river-specific CLs and assessment protocols and developments are reported annually to WGNAS (e.g. ICES, 2013).

The risk assessment frameworks applied by WGNAS directly evaluate the risk of meeting or exceeding the stock complex objectives. Managers can choose the risk level which they consider appropriate. ICES considers however that to be consistent with the MSY and the precautionary approach, and given that the CLs are considered to be limit reference points and to be avoided with a high probability, then managers should choose a risk level that results in a low chance of failing to meet the CLs. ICES

recommends that the probability of meeting or exceeding CLs for individual stocks should be greater than 95% (ICES, 2012b).

3.1.2 Reference points in the NEAC area

River-specific CLs have been derived for salmon stocks in some countries in the NEAC area (France, Ireland, UK (England & Wales) and Norway). An interim approach has been developed for estimating national CLs for countries that cannot provide one based upon river-specific estimates. This approach is based on the establishment of pseudo stock–recruitment relationships for national salmon stocks (Potter *et al.*, 2004).

The NEAC-PFA run reconstruction model (see below) provides a means of relating estimates of the numbers of recruits to the numbers of spawners. The numbers of 1SW and MSW spawners are converted into numbers of eggs deposited using the proportion of female fish in each age class and the average number of eggs produced per female. The egg deposition in year ‘n’ is assumed to contribute to the recruitment in years “n+3” to “n+8” in proportion to the numbers of smolts produced of ages 1 to 6 years respectively. These proportions are then used to estimate the ‘lagged egg deposition’ contributing to the recruitment of maturing and non-maturing 1SW fish in the appropriate years. The plots of lagged eggs (stock) against the 1SW adults in the sea (recruits) are presented as ‘pseudo stock–recruitment’ relationships for each homewater country or region that is unable to provide river-specific CLs. In countries where with more than one region, the analysis is carried out for each region separately and the resulting estimates are summed to provide a national figure.

As noted previously, ICES currently define the CL for salmon as the stock size that will result in the maximum sustainable yield (MSY) in the long term. However, it is not straightforward to estimate this point on the stock–recruitment relationships established by the national PFA run-reconstruction models, as the replacement line (i.e. the line on which ‘stock’ equals ‘recruits’) is not known for these relationships. This is because the stock is expressed as eggs, while the recruits are expressed as adult salmon. To address this, WGNAS has developed a method for setting biological reference points from the national/ regional pseudo stock–recruitment datasets (ICES, 2001). This model assumes that there is a critical spawning stock level below which recruitment decreases linearly towards zero and above which recruitment remains constant. The position of this critical stock level is determined by searching for the stock value that provides the line of best fit for the stock and recruitment data provided by the PFA run-reconstruction model as determined by the residual sum of squares. This point is a proxy for S_{lim} and is therefore defined as the CL for the stock, and is indicated by the inflection point in the hockey-stick relationship (e.g. see example at Figure 3.1.2.1).

Where river-specific estimates of CLs have been derived for all the rivers in a country or region, these are aggregated to provide national estimates. For countries where the development of river-specific CLs has not been completed, the method described above has been used (see example in Table 3.1.2.1, from ICES, 2013). The estimated national CLs are then summed to provide aggregate CLs for the northern and southern NEAC stock complexes (Table 3.1.2.1).

Table 3.1.2.1. Conservation limits (CLs) for NEAC countries and stock complexes estimated from river-specific values, where available, or the national PFA run- reconstruction model. Spawner escapement reserves (SERs) are also included for each stock complex.

WGNAS considers the current CL and SER levels may be less appropriate to evaluating the historical status of stocks (e.g. pre-1985), that in many cases have been estimated with less precision.

3.1.3 Reference points in the NAC area

In many regions of North America, the CLs are calculated as the number of spawners required to fully seed the wetted area of the river. The methods and values used to derive the egg and spawner conservation requirements for Atlantic Canada are documented in O'Connell *et al.* (1997). CLs have generally been derived using freshwater production dynamics translated to adult returns to estimate the spawning stock for maximum sustainable yield (MSY). Data were available for a limited number of stocks and these values were transported to the remaining rivers using information on habitat area and the age composition of the spawners. A similar procedure was used to determine the CLs for rivers in the USA (ICES, 1995). In Québec, adult-to-adult stock-recruitment relationships for six rivers were used to define the CLs for the other rivers (Caron *et al.*, 1999). The definition of conservation in Canada varies by region and in some areas, historically the values used were equivalent to maximizing /optimizing freshwater production. These are used in Canada as limit reference points and they do not correspond to MSY values. Reference points for Atlantic salmon are currently being reviewed for conformity with the Precautionary Approach policy in Canada. Revised reference points are expected to be developed.

The NAC conservation requirements for 2SW salmon (only these are required in developing catch options for the West Greenland fishery) are summarised in Table 3.1.3.1 (from ICES, 2013). These are calculated from the adult age structure within the different regions and total 123 349 2SW salmon for Canada and 29 199 2SW salmon for the USA, for a combined total of 152 548.

Table 3.1.3.1. 2SW Conservation limits (CLs) for the six regions in the NAC area estimated from river-specific values.

| COUNTRY AND COMMISSION AREA | STOCK AREA | 2SW SPAWNER REQUIREMENT |
|-----------------------------|---------------------|-------------------------|
| | Labrador | 34 746 |
| | Newfoundland | 4022 |
| | Gulf of St Lawrence | 30 430 |
| | Québec | 29 446 |
| | Scotia-Fundy | 24 705 |
| Canada Total | | 123 349 |
| USA | | 29 199 |
| North American Total | | 152 548 |

3.2 Estimating PFA

Estimates of PFA are derived by run-reconstruction methods. These work back in time from estimates of abundance in homewaters to earlier periods of the salmon's life cycle by adding in catches at appropriate times and adjusting for survival. The run-reconstruction approach was first presented at ICES in 1992 and was subsequently adopted for stocks on both sides of the Atlantic (Rago *et al.*, 1993a; Potter and Dunkley, 1993; Potter *et al.*, 1998; 2004). The main advantage of backwards-running, run-reconstruction models over alternative forward-running approaches is that more extensive data are available on adult returns (e.g. traps, counters and catch data) than on freshwater production of juveniles. In addition, rates of natural mortality (M) were thought to be lower and more stable for large salmon after their first winter in the sea than during the post-smolt phase (Potter *et al.*, 2003).

The models used to estimate PFA take the generalised form:

$$PFA = Nh * \exp(Mt_h) + \sum_i C_i * \exp(Mt_i)$$

Where: Nh is the number of adult fish returning to homewaters, C_i the catch of fish from the stock in each interception fishery i (operating before the fish return to homewaters), M the monthly instantaneous rate of natural mortality of salmon in the sea after the first sea-winter, t_i the time in months between the PFA date and the midpoint of fishery i, and t_h is the time in months between the PFA date and the midpoint of the return of fish to homewaters. Coastal catches are also added to the estimate where appropriate.

3.2.1 NEAC area run reconstruction model

The original model used to estimate the PFA of salmon from countries in the NEAC area was described by Potter *et al.* (2004); modifications have been described in subsequent WGNAS reports. PFA in the NEAC area is defined as the number of 1SW recruits on January 1st in the first sea winter. As there are relatively few fish of sea age three or more in most stocks, the model caters for two age groups, 1SW and MSW, the latter including all fish of sea age two or more that are treated as a single cohort. The model is therefore based on the annual catch in numbers of 1SW and MSW salmon in each country. These are raised to take account of minimum and maximum estimates of non-reported catches and exploitation rates of these two sea-age groups.

Thus, for each country (or region) c in year y, the total number of fish of sea age a caught in homewater fisheries (Ch_{a,y,c}) is calculated by dividing the declared catch (Cd_{a,y,c}) by the non-reporting rate (1 - U_{a,y,c}):

$$Ch_{a,y,c} = Cd_{a,y,c} / (1 - U_{a,y,c})$$

where: U_{a,y,c} is the estimated proportion of the total catch that is unreported or discarded. The number of fish returning to homewaters (Nh_{a,y,c}) is estimated by dividing the total homewater catch by the exploitation rate (H_{a,y,c}):

$$Nh_{a,y,c} = Ch_{a,y,c} / H_{a,y,c}$$

As the model provides estimates of total returns and total catch (including non-catch fishing mortality), it is then also possible to estimate the spawner escapement (Ns_{a,y,c}):

$$Ns_{a,y,c} = Nh_{a,y,c} - Ch_{a,y,c}$$

Total catches in the Faroese (Cf_{a,y}) and West Greenland (Cg_{a,y}) fisheries are similarly calculated by correcting the declared catches for non-reporting, but they are not raised for the exploitation rate, because the uncaught fish are accounted for from the returns to homewaters. The West Greenland fishery only exploits salmon that would otherwise mature as MSW fish, although the majority are 1SW fish in the summer that they are caught; for the purpose of the model, all are classed as 1SW. The Faroese fishery exploits predominantly MSW salmon, but also a small number of 1SW fish, 78% of which have been estimated to be maturing (ICES, 1994). Over the past two decades, a substantial proportion of the fish caught in the Faroese fishery have been escapees from salmon farms, and these are discounted from the assessment of wild stocks on the basis of data from Hansen *et al.* (1999). The incidence of farm escapees in the West Greenland

catch is thought to be <1.5% (Hansen *et al.*, 1997), so this portion is ignored in the model. The total estimated catches of wild fish in both distant-water fisheries are assigned to the PFA for different countries on the basis of historic tagging studies (Potter, 1996).

The returns to homewaters and catches in the distant-water fisheries of 1SW and MSW salmon are then raised to take account of the marine mortality between January 1st in the first sea winter (the PFA date) and the mid-point of the period over which the respective national fisheries operate. WGNAS determined a natural mortality value of 0.03 per month to be the most appropriate (ICES, 2002) and a range 0.02 to 0.04 is applied within the model in a Monte Carlo simulation. Thus, the PFA of maturing 1SW fish (PFAM), survivors of which will return to homewaters as 1SW adults, is:

$$PFAM_{y,c} = Nh_{1,y,c} * \exp(Mt_{h,1,c}) + 0.78 * Cf_{1,y} * w_y * pf_{1,c} * \exp(Mt_{f,1,c})$$

and the PFA of non-maturing 1SW fish (PFAN), survivors of which will return to homewaters as MSW adults, is:

$$PFAN_{y,c} = Nh_{2,y+1,c} * \exp(Mt_{h,2,c}) + Cg_{1,y} * pg_{1,c} * \exp(Mt_{g,1,c}) \\ + 0.22 * Cf_{1,y} * w_y * pf_{1,c} * \exp(Mt_{f,1,c}) + Cf_{2,y+1} * w_{y+1} * pf_{2,c} * \exp(Mt_{f,2,c})$$

where indices y and c represent year and country/region, indices 1 and 2 the 1SW and MSW sea age groups, w is the proportion of the Faroese catch that is of wild origin, pf and pg are the proportion of the catches in the Faroese and West Greenland fisheries originating in each country (as indexed), and t_h , t_f and t_g are the times in months between the PFA date and the midpoints of the homewater fisheries, the Faroese fishery, and the West Greenland fishery, respectively, for the year classes and country/region as indexed.

Total 1SW recruitment for the NEAC area in year y is therefore the sum of the maturing 1SW and non-maturing 1SW recruitments for that year for all countries:

$$PFA_y = \sum_c PFAM_{y,c} + \sum_c PFAN_{y,c}$$

The non-reporting rates, exploitation rates, natural mortality, and migration times in the above equations cannot be estimated precisely, so national experts provide minimum and maximum values based upon best available knowledge that are considered likely to be centred on the true values (ICES, 2003). A Monte Carlo simulation (12 000 trials) is used to estimate confidence intervals on the stock estimates.

Where appropriate to the provision of management advice, the national outputs from the model are combined into stock complexes, such as those for southern and northern NEAC (ICES, 2002). The confidence limits for these combined estimates are derived from the sum of the national variances obtained from the MCS (the covariances are assumed to be small). This model has provided time-series of PFA estimates for NEAC salmon stocks from 1971 to the present.

The model was initially run using 'Crystal Ball' (CB) in Excel (Decisioneering, 1996). However, an updated version of the model which runs in the 'R' programming language (R Development Core Team, 2007) was developed in 2011 (ICES, 2011). This provided a more flexible platform for the further development of the model and to allow its integration with the Bayesian forecast model for the development of catch options (see below). In 2012, the outputs of the CB and 'R' models were compared to examine the approaches taken and to validate the outputs (ICES, 2012a). Since 2013,

the run-reconstruction analysis has been completed by WGNAS using the 'R' programme (ICES, 2013). This has also enabled additional sources of uncertainty to be incorporated into the modelling approach (ICES, 2013).

The full set of country-specific data inputs, as used in the most recent assessment (ICES, 2015) is provided at Appendix 3. The 'R' code used for running the model and the additional data input file required to run the model, are available on the ICES WGNAS SharePoint site.

3.2.2 NAC area run reconstruction model

The run-reconstruction model developed by Rago *et al.* (1993a) and described in previous WGNAS reports (ICES, 2008; 2009a) and in the primary literature (Chaput *et al.*, 2005) is used to estimate returns and spawners by size (small salmon, large salmon) and sea age group (2SW salmon) to the six geographic regions of NAC from 1971 to the present. The model takes the form:

$$PFA_{year(i)} = [NR2_{year(i+1)} * e^{MX1} + NC2_{year(i+1)}] * e^{MX10} + NC1_{year(i)} + NG1_{year(i)}$$

where: $NR2_{year(i+1)}$ is the sum of 2SW returns to six regions of North America in year $i + 1$, $NC2_{year(i+1)}$ is the catch of 2SW salmon in Newfoundland and Labrador commercial fisheries in year $i + 1$, $NC1_{year(i)}$ is the catch of 1SW non-maturing salmon in Newfoundland and Labrador commercial fisheries in year i , $NG1_{year(i)}$ is the catch of 1SW non-maturing salmon of North American origin in the Greenland fishery in year i , and M is the monthly instantaneous natural mortality of 0.03.

The reconstruction begins with the estimation of returns of 2SW salmon in year $i + 1$ to six regions in eastern North America: Labrador, Newfoundland, Québec, Gulf, Scotia-Fundy, and USA. For the four southern regions, the regional returns include the harvest in the coastal commercial fisheries but this is not the case for Newfoundland and Labrador. For Labrador, the returns to rivers are estimated from the commercial harvest factored by an exploitation rate. The harvest of 2SW salmon in the Newfoundland and Labrador mixed-stock fisheries in year $i + 1$ is added to the sum of the returns to the six regions (prorated backward for one month of natural mortality - equates to 1st June of year $i + 1$) to produce the returns to North America. Finally, the harvests of North American origin salmon in the Greenland fisheries in year i and the harvest of non-maturing 1SW salmon in the Newfoundland and Labrador commercial fisheries in year i are added to the prorated returns to North America (ten months between abundance at Greenland on 1st August year i and North America on 1st June year $i + 1$) to produce the pre-fishery abundance of non-maturing 1SW salmon of North American origin. An instantaneous natural mortality rate of 0.03 per month is assumed for salmon in the second year at sea for all years (ICES, 2002). Adjustments to the input data resulting from reductions and subsequent closures of commercial fisheries in North America are summarized by Friedland *et al.* (2003).

Following earlier WGNAS recommendations (ICES, 2008), the run-reconstruction model since 2009 has been developed using Monte Carlo simulation (OpenBUGS; <http://mathstat.helsinki.fi/openbugs/>; Lunn *et al.*, 2000). This is similar to the approach applied for the NEAC area.

The PFA of the non-maturing component of 1SW fish, destined to be 2SW returns (excluding 3SW and previous spawners) is the estimated number of salmon in the North Atlantic on August 1st of the second summer at sea. As this requires estimates

of 2SW returns to rivers, there is always a lag in providing this figure (PFA estimates for year n require 2SW returns to rivers in North America in year $n + 1$).

The full set of data inputs, as used in the most recent assessment (ICES, 2015) is provided at Appendix 4. The 'R' code used for running the model is available on the WGNAS SharePoint site.

3.2.3 Instantaneous natural mortality rate (M)

The natural mortality rate for salmon after they recruit to the distant water fisheries has been the subject of much discussion. WGNAS originally used a value of 0.01 per month, based upon Doubleday *et al.* (1979), but this was modified to 0.03 per month following a detailed review as part of the EU SALMODEL project (Crozier *et al.*, 2003; ICES, 2002) on the basis of inverse-weight and maturity-schedule models. The rate is assumed to have been constant over the time-series. While mortality may be expected to vary among years and may also be different for maturing and non-maturing 1SW recruits, WGNAS has not had data on which to base the use of different values, or values that change over time. The assumption is, therefore, that the mortality of adult fish after the first winter at sea has not changed and that all the variability of marine mortality has occurred at the post-smolt stage. Efforts are continuing to explore levels of mortality and to better partition this between different stages of the life cycle. The issue was also subject to further investigation within the EU ECOKNOWS project and Bayesian modelling may provide alternative approaches in future.

3.3 PFA forecast models

3.3.1 Introduction

The provision of quantitative catch advice for the distant water fisheries requires estimates of abundance before the fisheries take place. While there has been some use of in-season surveys in the management of these fisheries (NASCO, 2001), such methods are considered too impractical and costly to implement on a widespread scale (Potter *et al.*, 2004). Models have therefore been developed by WGNAS which relate abundance estimates obtained at other life stages to the PFA. The objective has been to account for this relationship in terms of biological or environmental factors that affect natural mortality, and to use this to forecast future stock levels.

An initial PFA forecast model for North American stocks (Rago *et al.*, 1993b) utilised indices of thermal habitat in relation to historically observed PFA (from the run-reconstruction model) to predict future PFA. Similar approaches were explored by Crozier *et al.* (2003) for NEAC stocks. However, while statistically significant temperature indices could be constructed, the relationships were not always consistent or intuitively correct. Alternative approaches were therefore explored for NEAC; these are described by Potter *et al.* (2004). More recently work by the ICES Study Group on Salmon Stock Assessment and Forecasting (SGSSAFE) has, however, resulted in the development of Bayesian forecast models for both NAC and NEAC (ICES, 2009a; 2011; Chaput, 2012).

In the latest models, PFA dynamics by complex are modelled using the estimates of adult spawners, adjusted to the number of eggs per fish based on life-history characteristics of the age groups within each region of the stock complexes (ICES, 2011; Chaput, 2012). The spawner to PFA dynamic is modelled as:

$$PFA_y = e^{\alpha_y} LE_y e^{\epsilon}$$

where: α_y is the productivity parameter from eggs ($\times 1000$) to PFA (number of fish) for PFA year y (on a log-scale), LE_y the estimated lagged eggs ($\times 1000$) corresponding to the PFA cohort in year y , and the progress of α_y is modelled as $\alpha_{y+1} = \alpha_y + \varepsilon$, with $\varepsilon \sim N(0, \sigma^2)$.

Productivity is modelled as an integration of survival in freshwater and during the first year at sea. An important assumption is the absence of heritability of age at maturity, i.e. all eggs are considered equivalent regardless of the age of the spawners. Lagged eggs refer to the adjustment of the egg depositions to correspond to the expected age at smoltification. Spawners in year 'n' contribute to recruitment in years 'n+3' to 'n+8' depending upon the relative proportions of one to six year-old smolts that they produce. For example, spawners in year 'n' produce eggs that hatch in year 'n+1' and may produce one year-old smolts in year 'n+2', which would become 1SW recruits in year 'n+3'. Any two year-old smolts from the same spawners would produce 1SW recruits in year 'n+4', etc.

At the stock complex level, lagged eggs are the sum of the eggs from the spawners in year $y - (s + 2)$ weighted by the proportion of the smolts produced at age s in region k summed over regions in the complex. Two years are added to the smolt age, for the spawning year and smolt migration year, to lag the eggs to the corresponding year of PFA:

$$LE_y = \sum_k \sum_s Eggs_{y-(s+2),k} * prop_{s,k}$$

3.3.2 NEAC PFA Forecast model

A forecast model to estimate PFA for all four NEAC stock complexes has been developed in a Bayesian framework by the Study Group on Salmon Stock Assessment and Forecasting (SGSSAFE). The model was originally reported to WGNAS in 2009 (ICES, 2009a), but was subsequently refined and has been in use by WGNAS in its present form since 2011 (ICES, 2011). The models for the northern and southern NEAC stock complexes have exactly the same structure and are run independently. A Directed Acyclical Graph (DAG) for the models is provided in Figure 3.3.2.1. The model considers both the maturing PFA (denoted PFA_m) and the non-maturing PFA (denoted PFA_{nm}). The full code used for running the model is available on the WGNAS SharePoint site.

A disaggregated version of the Bayesian model has since been developed using the same structure to provide forecasts at a country level, for all countries in both southern and northern NEAC model implementations (ICES, 2013). In these, countries are linked hierarchically only through the variance on the productivity parameter "a". There is no modelling linkage between the northern and southern complexes.

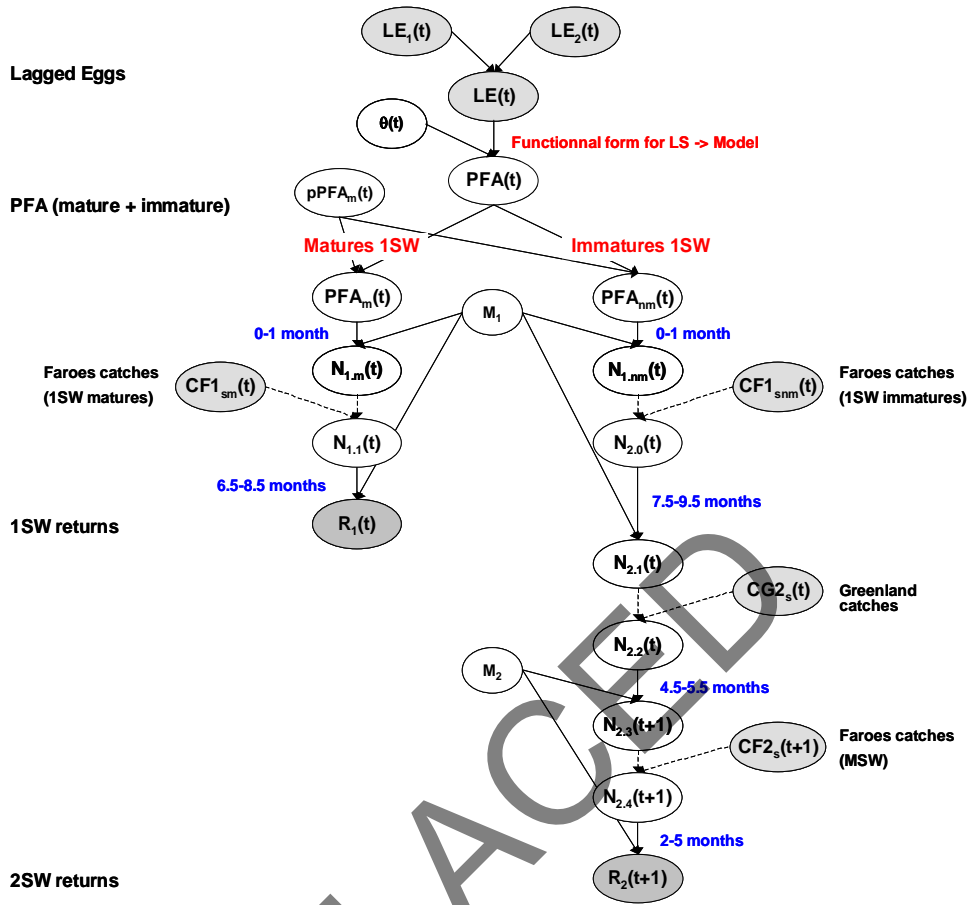


Figure 3.3.2.1. Directed Acyclical Graph (DAG) of the structure of the combined sea age model for the southern NEAC and northern NEAC forecast models. Ellipses in grey are observations (or pseudo-observations) derived from sampling programmes or from submodels (run-reconstruction).

The PFA is modelled using the summation of lagged eggs from 1SW and MSW fish (LE) for each year t and an exponential productivity parameter (a).

$$PFA_t = LE_t^* \exp(at)$$

The productivity parameter (a) is the proportionality coefficient between lagged eggs and PFA . This is forecasted one year at a time (a_{t+1}) in an auto correlated random walk, using the previous year's value (a) as the mean value in a normal distribution, with a common variance for the time-series of a .

$$a_{t+1} = a_t + \varepsilon_t \quad \varepsilon_t \sim N(0, a.\sigma^2)$$

The maturing PFA (denoted PFA_m) and the non-maturing PFA (denoted PFA_{nm}) recruitment streams are subsequently calculated from the proportion of PFA maturing ($p.PFA_m$) for each year t . $p.PFA_m$ is forecast as an autocorrelated value from a normal distribution based on a logit scale, using the previous year's value as the mean and a common variance across the time-series of $p.PFA_m$.

$$\begin{aligned} \text{logit}.p.PFA_{mt+1} &\sim N(\text{logit}.p.PFA_{mt}, p.\sigma^2) \\ \text{logit}.p.PFA_{mt} &= \text{logit}(p.PFA_{mt}) \end{aligned}$$

Uncertainties in the lagged eggs were accounted for by assuming that the lagged eggs of 1SW and MSW fish were normally distributed with means and standard deviations

derived from the Monte-Carlo run-reconstruction at the scale of the stock complex. The uncertainties in the maturing and non-maturing PFA returns are derived in the Bayesian forecast models through the pseudo-observation method proposed by Michielsens *et al.* (2008), as used in the NAC model.

The natural mortality in the post-PFA time point was assumed constant among years, centred on an instantaneous rate value of 0.03 per month with a 95% confidence interval range of 0.02 to 0.04.

Catches of salmon at sea in the West Greenland fishery (as 1SW non-maturing salmon) and at Faroes (as 1SW maturing and MSW salmon) were introduced as covariates and incorporated directly within the inference and forecast structure of the model. For southern NEAC, the data are available for a time-series of lagged eggs and returns commencing in 1978. Although the return estimates to southern NEAC begin in 1971, the lagged eggs are only available from 1978 due to the smolt age distributions (one to five years). For northern NEAC, data are available for a shorter time-series. Return and spawner estimates begin in 1983, but due to the smolt age distributions (one to six years), the lagged eggs are only available from 1991 onward. The models are fitted and forecasts derived in a consistent Bayesian framework.

The model provides forecasts for maturing and non-maturing stocks for both southern and northern NEAC complexes (and countries) for five years. Risks are defined each year as the posterior probability that the PFA would be greater than or equal to the age and stock complex/ country specific Spawner Escapement Reserves (SERs), under the scenario of no exploitation.

The country disaggregated version of the Bayesian NEAC inference and forecast model incorporates country specific catch proportions at Faroes, lagged eggs and returns of maturing and non-maturing components. Model structure and operation is as described above, incorporating country and year indexing. Linkage between countries in the model is through the common variance parameter associated with the productivity parameter (a) (the proportionality coefficient between lagged eggs and PFA), which is forecast forward and used along with the forecast proportion maturing to estimate the future maturing and non-maturing PFAs. The evolution of a is independent between countries with the exception of its associated variance. Evolution of the proportion maturing ($p.PFAm$) is also independent for each country, as is its variance.

3.3.3 NAC PFA Forecast model

WGNAS (ICES, 2009; 2012a) developed forecasts of the pre-fishery abundance for the non-maturing 1SW salmon (PFA) using a Bayesian framework that incorporates the estimates of lagged spawners and works through the fisheries at sea to determine the corresponding returns of 2SW salmon, conditioned by fisheries removals and natural mortality at sea. This model considers regionally-disaggregated lagged spawners and returns of 2SW salmon for the six regions of North America. The model is summarised in the Directed Acyclical Graph in Figure 3.3.1.1. The year is identified by the i index. The full code used for running the model is available on the WGNAS SharePoint site.

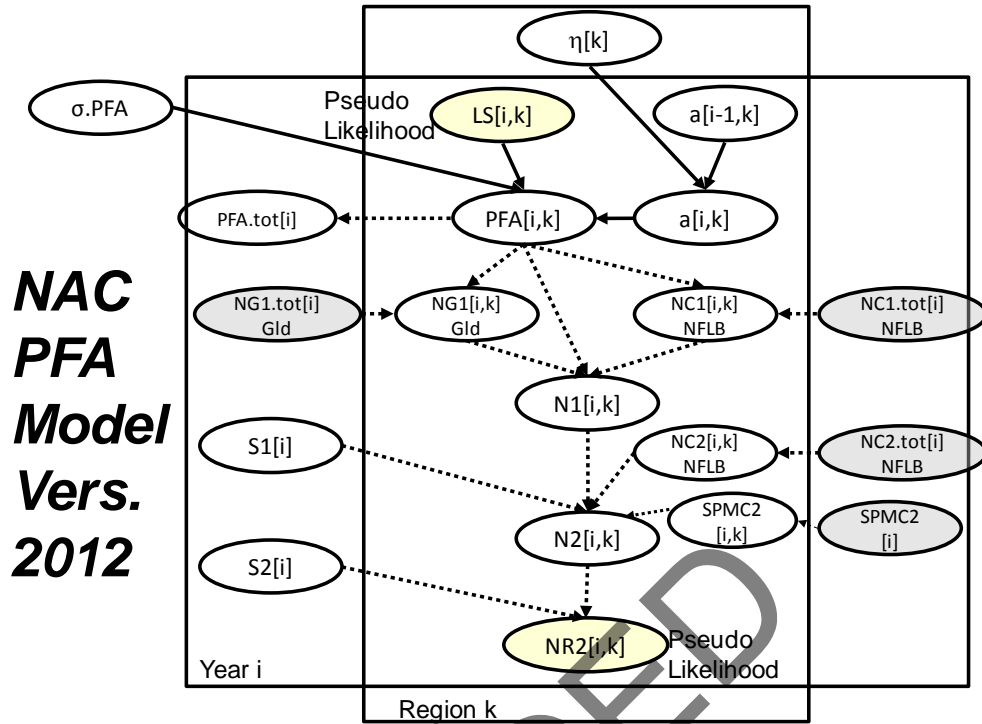


Figure 3.3.1.1. Directed Acyclical Graph (DAG) of the structure of the region disaggregated forecast model for 2SW salmon of North American origin. Ellipses in grey are observations (or pseudo-observations) derived from sampling programmes or from submodels (run-reconstruction).

Lagged spawners $LS_{i,k}$ represent the sum of smolt age adjusted annual spawners by region (k) that would be expected to contribute to the recruitment at sea prior to the fisheries (PFA) for year i . $LS_{i,k}$ are not directly observed but are estimated from the run-reconstruction submodel used to estimate returns and spawners to each of the six regions.

The probability distributions of LS (and returns of 2SW) by region are used as likelihood functions expressing comparative degrees of belief given the data and a probability model not explicitly specified in the current model. The probability distributions were drawn from the Monte Carlo simulations and assumed to be normal with known mean ($LS.m$) and precision ($1/\text{variance}$) ($\text{tau}.LS$). The use of this distribution as a likelihood function is equivalent to assuming a pseudo-observation equal to $LS.m$ issuing from a sampling distribution with mean and precision equal to LS and $\text{tau}.LS$ (Michielsens *et al.*, 2008).

$$LS.m_{i,k} \sim N(LS_{i,k}, \text{tau}.LS_{i,k})$$

The $LS.m_{i,k}$ (mean) and $\text{tau}.LS_{i,k}$ (precision) were derived assuming the lagged spawner values issued from a normal distribution characterized by the 95% confidence interval range statistics retained from the Monte Carlo simulations of returns.

Similarly, the returns of 2SW salmon to the six regions ($NR2_{i,k}$) are not directly observed, but estimated from the run-reconstruction model. The probability distributions were assumed to be normal with known mean $NR2.m$ and variance $\text{tau}.NR2$. As with the LS variable, the $NR2$ were treated as pseudo-observations equal to $NR2.m$ issuing from normal sampling distributions with means and variances equal to $NR2$ and $\text{tau}.NR2$.

$$NR2.m_{i,k} \sim N(NR2_{i,k}, \tau.NR2_{i,k})$$

In between the lagged spawners and returns as 2SW salmon, the catches in the various sea fisheries and conditioning for natural mortality as the fish move from the time of the PFA to homewaters, are incorporated (Figure 3.3.1.1). The catches in the commercial fisheries of West Greenland and the Newfoundland and Labrador commercial and coastal fisheries (NG1.tot, NC1.tot and NC2.tot) are not directly observed, but estimated with error. The catches are converted to numbers of fish of 1SW non-maturing and 2SW fish based on the characteristics of the fish in the catch. Their (prior) probability distributions are obtained from catch statistics according to a formal structure included in the model.

Catches of large salmon (assumed to be 2SW salmon) from the St Pierre & Miquelon fisheries (SPMC) are also included in the model as point estimates.

The natural mortality in the post-PFA time point was assumed constant between years, centred on an instantaneous rate value of 0.03 per month (95% confidence interval range of 0.02 to 0.04).

For the NAC 2SW component, the model is fitted to an historical dataseries of lagged eggs starting from 1978. Although the return and spawner estimates for NAC begin in 1971, the lagged eggs are only available from 1978 due to the smolt age distributions (one to six years).

The years are modelled independently conditionally on the lagged spawners and yearly productivity parameters. The lagged spawners to PFA ratios (productivity) are modelled dynamically, i.e. assuming they are sequentially dependent within a region and attempts to take into account the most significant sources of uncertainty. The DAG for the model is shown in Figure 2.

PFA is assumed to be proportional to lagged-spawners (LS), with i.i.d. lognormal errors, and is modelled separately for each region ($k = 6$). The first year in the time-series (t) is 1978 for lagged spawners (due to the range of smolt ages 1 to 6 for NAC and the start of the spawner time-series in 1970) and the last year of lagged spawner data is for the 2017 PFA year. The PFA can be modelled for 1978 to 2013 (the last PFA year for which returns of 2SW salmon have been estimated back to rivers in 2014).

$$PFA_{i,k} = \overline{LogN(PFA_{i,k}, \sigma.PFA^2)}$$

$$\overline{PFA_{i,k}} = \log(LS_{i,k}) + a_{i,k} + \varepsilon_{i,k}$$

$$\varepsilon_{i,k} \stackrel{iid}{\sim} N(0, \sigma.PFA^2)$$

The total PFA is calculated as the sum of the regional PFA's ($k = 6$). The proportion of the total PFA in each region is calculated directly as:

$$p.PFA_{i,k} = PFA_{i,k} / PFA.tot$$

A non-informative prior is assumed for $\sigma.PFA^2$ ($1/\sigma.PFA^2 \sim \text{gamma}(0.01, 0.01)$)

The proportionality coefficient (log) $a_{i,k}$ between $LS_{i,k}$ and $PFA_{i,k}$ for each region is modelled dynamically as a random walk with a year and region residual variation ($\eta_{i,k}$) assumed multivariate normal (MVN). The variance covariance matrix (Σ) allows for correlations among regional productivity values reflecting that the fish

share a common marine environment during part of their life cycle and that there are regional specificities in the evolution of the freshwater or the marine coastal environment.

$$a_{i+1,k} = a_{i,k} + \eta_{i,k}$$

$$\eta_{i,k} \stackrel{iid}{\sim} MVN(0, \Sigma)$$

The common yearly evolution of a_i is the mean of annual a across regions:

$$a.y_i \leftarrow \text{mean}(a_{i,k})$$

This parameterization of the covariance of the proportionality coefficient differs from an earlier version of the model for which the proportionality (log) coefficient $a_{i,k}$ between $LS_{i,k}$ and $PFA_{i,k}$ for each region dynamically as a random walk model with the addition of a regionally common annually varying parameter ($e.y_i$).

$$a_{i+1,k} = a_{i,k} + e.y_{i+1} + \omega_{i,k} \quad \text{with} \quad \omega_{i,k} \stackrel{iid}{\sim} N(0, a.\sigma_k^2)$$

$$e.y_i \stackrel{iid}{\sim} N(0, y.\sigma^2)$$

The correlation matrix of a among the regions is calculated from the covariance matrix:

- 1) the precision matrix is inverted to produce the covariance matrix;
- 2) the covariance matrix is transformed to the correlation matrix.

The positive-definite matrix (T , the precision matrix) is inverted:

$$\text{covariance matrix} \leftarrow \text{solve}(T)$$

$$\text{correlation matrix} \leftarrow \text{cov2cor}(b)$$

The dynamic component of the model requires initialization for the first year ($i = 1978$) and an uninformative prior is assumed:

$$a_{1,k} \stackrel{iid}{\sim} N(0, 100)$$

The models are fitted and forecasts derived in a consistent Bayesian framework under the OpenBUGS 3.0.3 software (<http://mathstat.helsinki.fi/openbugs/>; Lunn *et al.*, 2000).

3.3.4 Summary of NAC and NEAC forecast models

The data inputs and models currently used by WGNAS for forecasting and provision of catch advice differ between the Commission areas; outline details are summarised in the text table below.

| FORECAST MODELS | | |
|---------------------|---------|--------------------------------------------------------|
| | NAC | NEAC |
| Data inputs | | |
| Time period of data | 1978 on | 1978 on for southern NEAC 1991 on for northern NEAC |

| FORECAST MODELS | | |
|------------------------------|-----------------------------------------------------------------------------------|-------------------------------------------------------------------------------------------------------------------------------------------------------|
| | NAC | NEAC |
| Spatial aggregation | Separately for six regions of North America | By southern and northern stock complexes & NEAC countries |
| Age components | 2SW salmon component only | 1SW and MSW age components |
| Spawners | Lagged spawners by region for 2SW salmon only | Lagged eggs by sea age component for the southern and northern complexes/ country |
| Returns | Returns by region of 2SW salmon only | Returns of 1SW and MSW age components by stock complex / country |
| Model structure | | |
| Spatial aggregation | Spawners and returns of 2SW salmon for six regions | Spawners and returns for two sea age components for the southern and northern NEAC complexes / countries |
| Dynamic function | Random walk dynamic | Random walk dynamic |
| | Region-specific recruitment rates linked with an annual recruitment rate variable | Sea age specific recruitment rates linked with a probability of maturing variable |
| Latent variables of interest | PFA 1SW non-maturing Recruitment rate by region and year | PFA 1SW maturing and PFA 1SW non-maturing by stock complex/ country Recruitment rate by sea age component and the probability of maturing variable |
| Forecast years | Four years | Five years – i.e. the present year - 1, the present year, and the next 3 years (y-1 is a forecast, as the MSW stock component is yet to return). |

3.4 The development of a risk analysis framework for catch advice

3.4.1 Introduction

The provision of catch advice in a risk framework involves incorporating the uncertainty in all the factors used to develop the catch options (ICES, 2002). The ranges in the uncertainties of all the factors will result in assessments of differing levels of precision. The analysis of risk involves four steps:

- 1) identifying the sources of uncertainty;
- 2) describing the precision or imprecision of the assessment;
- 3) defining a management strategy; and
- 4) evaluating the probability of an event (either desirable or undesirable) resulting from the fishery action.

The uncertainties have been identified and quantified in the assessment of PFA for salmon stocks in both the NAC and NEAC areas. NASCO's strategy for the management of salmon fisheries is based upon the principle of ensuring that stocks are above CLs (defined in terms of spawner escapement or egg deposition) with a high

probability. The undesirable event to be avoided is that the spawning escapement after the fisheries will be below the CLs.

3.4.2 Catch advice and risk analysis framework for the West Greenland fishery

A risk framework for the provision of catch advice for the West Greenland fishery has been applied since 2003 (ICES, 2003) and has been subject to a number of subsequent updates. The current procedure is outlined below. This involves estimating the uncertainty in meeting defined management objectives at different levels of catch (catch options). The risk framework has been formally accepted by NASCO.

Two stock complexes are of relevance to the management of the West Greenland fishery; non-maturing 1SW fish from North America and non-maturing 1SW fish from southern NEAC. The risk assessments for the two stock complexes are developed in parallel and then combined at the end of the process into a single summary plot or catch options table. The risk analysis proceeds as illustrated in the flowchart in Figure 3.4.2.1).

The primary inputs to the risk analysis for the complex at West Greenland are:

- PFA forecast for the year of the fishery; PF_{ANA} and PF_{NEAC} ;
- Harvest level being considered (t of salmon);
- Conservation spawning limits or alternate management objectives; and
- The post-fishery returns to each region.

The risk analysis of catch options incorporates the following input parameter uncertainties: (i) the uncertainty of the pre-fishery abundance forecast, (ii) the uncertainty in the biological parameters used to translate catches (weight) into numbers of salmon, and (iii) the uncertainty in attaining the conservation requirements simultaneously in different regions.

The uncertainty in the PF_{ANA} and PF_{NEAC} is accounted for in the forecast approach described above. The number of 1SW non-maturing fish of North American and European origin in a given catch (t) is conditioned by the continent of origin of the fish ($prop_{NA}$, $prop_E$), by the average weight of the fish in the fishery ($W_{tAllages}$), and by the proportion 1SW non-maturing fish in the respective continent of origin catches. These parameters define how many fish originating in North America and Europe are expected in the fishery harvests. For a level of fishery under consideration, the weight of the catch is converted to number of fish of each continent of origin using the following equation:

$$C1SW_c = \frac{t \times prop_C}{ACF * (prop_{NA} \times W_{t1SW_{NA}} + prop_E \times W_{t1SW_E})}$$

where: $C1SW_c$ is the catch (number of fish) of 1SW salmon originating in continent C (either North America or Europe), t is the fishery harvest at West Greenland in kg, $prop_C$ is the proportion of the 1SW salmon harvest which originates from continent C , $W_{t1SW_{NA}}$ and W_{t1SW_E} are the average weight (kg) in the fishery of a 1SW salmon of North American and European origin, respectively, and ACF is the age correction factor by weight for salmon in the fishery which are not at age 1SW.

Since these parameters are not known for the forecast years of interest, they are estimated from previous values. Thus, $prop_{NA}$ (and $prop_{NEAC}$ as $1 - prop_{NA}$) are drawn randomly from observed values of the past five years taking account of uncertainty

due to sample sizes. For the other parameters, it is assumed that the parameters for $W_{tAllages}$ and the proportion non-maturing 1SW in the catch by continent of origin could vary uniformly within the values observed in the past five years.

For a level of fishery under consideration, the weight of the catch is converted to fish of each continent of origin and subtracted from one of the simulated forecast values of PFA_{NA} and PFA_{NEAC} . The fish that escape the Greenland fishery are immediately discounted by the fixed sharing fraction (F_{na}) historically used in the negotiations of the West Greenland fishery. The sharing fraction chosen is the 40%:60% West Greenland:North America split. The same sharing arrangement has been used for NEAC stocks (ICES, 2003). [Any sharing fraction could be considered and incorporated at this stage of the risk assessment].

After the fishery, fish returning to homewaters are discounted for natural mortality from the time they leave West Greenland to the time they return to rivers. For North America this is a total of eleven months at a rate of $M = 0.03$ (equates to 28.1% mortality). For Southern NEAC stocks this is a total of eight months at a rate of $M = 0.03$ (equates to 21% mortality). The fish that survive to North American homewaters are then distributed among the regions based on the regional proportions of lagged spawners for the last five years when estimates of spawners were available. The uncertainty in the regional proportions is characterised by drawing at random from a uniform distribution defined by the minimum and maximum regional ranges from the five years and calculating the average proportion for each of the six regions in North America.

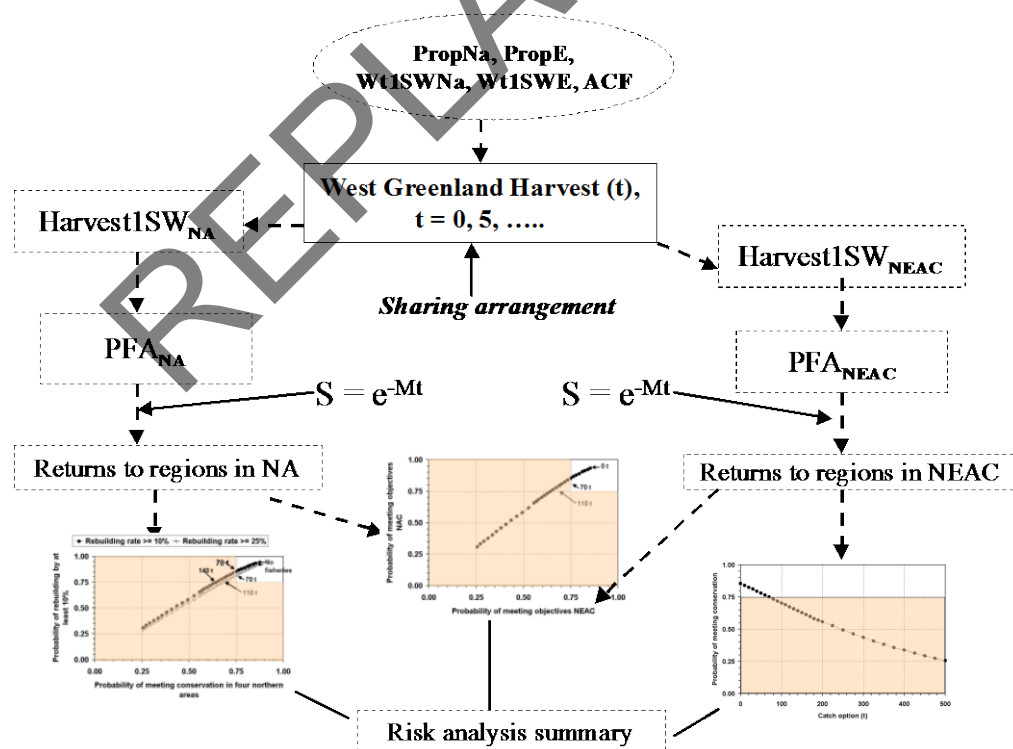


Figure 3.4.2.1. Flow chart summarising risk analysis for catch options at West Greenland using the PFA_{NA} and the PFA_{NEAC} predictions for the year of the fishery. Inputs with solid borders are considered known without error. Estimated inputs with observation error that is incorporated in the analysis have dashed borders. Solid arrows are functions that introduce or transfer without error whereas dashed arrows transfer errors through the components.

The final step in the risk analysis of the catch options involves combining the conservation requirement or alternate management objectives with the probability distribution of the returns to North America for different catch options. Estimated 2SW returns to each region are compared to the conservation objectives of Labrador, Newfoundland, Québec, and Gulf. Estimated returns for Scotia-Fundy are compared to the objective of achieving an increase of 25% relative to average returns of the base period, 1992–1996. For the USA, the management objective was revised in 2014 (ICES, 2014). Estimated returns for the USA are now compared to the objective of achieving 2SW adult returns of 4549 or greater. The advice to fisheries managers is presented as a probability plot (or table) of meeting or exceeding the objectives relative to increasing harvest levels at West Greenland.

ICES has adopted, a risk level of 75% of simultaneous attainment of seven management objectives (ICES, 2003) as part of an agreed management plan for the West Greenland fishery. The same level of risk aversion is applied for catch advice for homewater fisheries on the North American stock complex.

The catch advice for the West Greenland fishery is currently tabulated to show the probability of each management unit achieving its CL (or alternative reference level) individually and the probability of this being achieved by all management units simultaneously (i.e. in the same given year) (e.g. ICES, 2012a). This allows managers to evaluate both individual and simultaneous attainment levels in making their management decisions. Table 3.4.2.1 provides an example of catch options for West Greenland for the years 2012 to 2014 (ICES, 2012a). An updated catch options table for 2015 to 2017 is provided in Section 5.3 of the 2015 Working Group report (see above).

The models currently used by WGNAS in developing catch advice are considered to provide consistent characterisation of the status and expectations for Atlantic salmon in the North Atlantic. Compared to previous models used by WGNAS prior to 2009, the Bayesian models provide more flexibility, are consistent with the emerging emphasis on such approaches in natural resource assessment, and can provide management advice consistent with the probability of achieving management objectives.

Table 3.4.2.1. Catch options tables for mixed-stock fishery at West Greenland by year of PFA, 2012 to 2014.

| 2012 CATCH OPTION (T) | PROBABILITY OF MEETING OR EXCEEDING REGION-SPECIFIC MANAGEMENT OBJECTIVES | | | | | | | |
|-----------------------------|---------------------------------------------------------------------------|------|------|------|------|------|--------|------|
| | LAB | NFLD | QC | GULF | SF | USA | S-NEAC | ALL |
| 0 | 0.45 | 0.86 | 0.71 | 0.50 | 0.15 | 0.89 | 0.98 | 0.06 |
| 10 | 0.42 | 0.84 | 0.67 | 0.48 | 0.14 | 0.88 | 0.98 | 0.05 |
| 20 | 0.40 | 0.83 | 0.63 | 0.45 | 0.13 | 0.87 | 0.98 | 0.05 |
| 30 | 0.38 | 0.81 | 0.59 | 0.42 | 0.12 | 0.85 | 0.98 | 0.04 |
| 40 | 0.36 | 0.78 | 0.54 | 0.40 | 0.12 | 0.83 | 0.98 | 0.04 |
| 50 | 0.34 | 0.76 | 0.50 | 0.38 | 0.11 | 0.81 | 0.98 | 0.03 |
| 60 | 0.32 | 0.73 | 0.46 | 0.36 | 0.10 | 0.79 | 0.98 | 0.03 |
| 70 | 0.30 | 0.70 | 0.42 | 0.33 | 0.09 | 0.77 | 0.98 | 0.03 |
| 80 | 0.28 | 0.67 | 0.39 | 0.31 | 0.08 | 0.74 | 0.98 | 0.03 |
| 90 | 0.26 | 0.64 | 0.35 | 0.29 | 0.08 | 0.72 | 0.97 | 0.02 |
| 100 | 0.24 | 0.60 | 0.32 | 0.27 | 0.07 | 0.68 | 0.97 | 0.02 |
| 2013 CATCH OPTION (T) | PROBABILITY OF MEETING OR EXCEEDING REGION-SPECIFIC MANAGEMENT OBJECTIVES | | | | | | | |
| | LAB | NFLD | QC | GULF | SF | USA | S-NEAC | ALL |
| 0 | 0.48 | 0.78 | 0.73 | 0.50 | 0.25 | 0.75 | 0.95 | 0.08 |
| 10 | 0.46 | 0.76 | 0.70 | 0.48 | 0.24 | 0.73 | 0.95 | 0.07 |
| 20 | 0.44 | 0.75 | 0.67 | 0.46 | 0.23 | 0.72 | 0.95 | 0.06 |
| 30 | 0.42 | 0.73 | 0.63 | 0.44 | 0.22 | 0.70 | 0.95 | 0.06 |
| 40 | 0.41 | 0.70 | 0.60 | 0.42 | 0.21 | 0.68 | 0.95 | 0.06 |
| 50 | 0.39 | 0.68 | 0.56 | 0.40 | 0.20 | 0.66 | 0.94 | 0.05 |
| 60 | 0.37 | 0.65 | 0.53 | 0.38 | 0.19 | 0.64 | 0.94 | 0.05 |
| 70 | 0.35 | 0.63 | 0.50 | 0.36 | 0.18 | 0.62 | 0.94 | 0.05 |
| 80 | 0.33 | 0.60 | 0.47 | 0.34 | 0.17 | 0.59 | 0.94 | 0.04 |
| 90 | 0.31 | 0.57 | 0.44 | 0.32 | 0.16 | 0.57 | 0.94 | 0.04 |
| 100 | 0.30 | 0.54 | 0.41 | 0.31 | 0.15 | 0.55 | 0.94 | 0.04 |
| 2014 CATCH OPTION (T) | PROBABILITY OF MEETING OR EXCEEDING REGION-SPECIFIC MANAGEMENT OBJECTIVES | | | | | | | |
| | LAB | NFLD | QC | GULF | SF | USA | S-NEAC | ALL |
| 0 | 0.56 | 0.78 | 0.75 | 0.55 | 0.20 | 0.86 | 0.94 | 0.08 |
| 10 | 0.55 | 0.77 | 0.73 | 0.53 | 0.20 | 0.85 | 0.94 | 0.08 |
| 20 | 0.53 | 0.75 | 0.70 | 0.51 | 0.19 | 0.84 | 0.94 | 0.07 |
| 30 | 0.52 | 0.73 | 0.67 | 0.49 | 0.18 | 0.83 | 0.94 | 0.07 |
| 40 | 0.50 | 0.71 | 0.64 | 0.47 | 0.17 | 0.82 | 0.94 | 0.06 |
| 50 | 0.48 | 0.69 | 0.62 | 0.46 | 0.17 | 0.81 | 0.94 | 0.06 |
| 60 | 0.46 | 0.67 | 0.59 | 0.44 | 0.16 | 0.79 | 0.94 | 0.06 |
| 70 | 0.45 | 0.65 | 0.56 | 0.42 | 0.16 | 0.77 | 0.94 | 0.05 |
| 80 | 0.43 | 0.63 | 0.54 | 0.41 | 0.15 | 0.76 | 0.94 | 0.05 |
| 90 | 0.42 | 0.61 | 0.51 | 0.39 | 0.14 | 0.74 | 0.94 | 0.05 |
| 100 | 0.40 | 0.59 | 0.49 | 0.38 | 0.14 | 0.72 | 0.94 | 0.05 |

3.4.3 Catch advice and risk analysis framework for the Faroes fishery

3.4.3.1 Outline of the risk framework and management decisions required

There is currently no agreed framework for the provision of catch advice for the Faroes fishery adopted by NASCO. However, NASCO has asked ICES, for a number of years, to provide catch options or alternative management advice with an assessment of risks relative to the objective of exceeding stock conservation limits for salmon in the NEAC area. An initial risk framework that could be used to provide and evaluate catch options for the Faroes fishery was outlined by WGNAS in 2010 (ICES, 2010a). This was based on the method currently used to provide catch advice for the West Greenland fishery, which involves estimating the uncertainty in meeting defined management objectives at different catch levels (TAC options). The Faroes risk framework was developed further at subsequent WGNAS meetings (ICES, 2011; 2012a) and the current proposed procedure is outlined below.

A number of decisions are required by managers before full catch advice could be provided (ICES, 2011; 2012a). Specifically, ICES has indicated that NASCO would need to agree upon the following issues before the risk framework could be finalised:

- season (January to December or October to May) to which any TAC should apply;
- share arrangement for the Faroes fishery;
- choice of management units for NEAC stocks; and
- specification of management objectives.

In developing an indicative risk framework, WGNAS has made pragmatic choices regarding these issues:

Faroes fishing season: A decision is required on the period to which any TAC for the Faroes fishery would apply. The Faroes fishery has historically operated between October/November and May/June, but the historical TACs applied to a calendar year. This means that two different cohorts of salmon of each age class (e.g. two cohorts of 1SW salmon, etc.) were exploited under each TAC. ICES (2011) recommended that NASCO manage any fishery on the basis of fishing seasons operating from October to June, and catch advice should be provided on this basis.

Sharing agreement: The 'sharing agreement' establishes the proportion of any harvestable surplus within the NEAC area that could be made available to the Faroes fishery through the TAC. In effect, for any TAC option being evaluated for the Faroes, it is assumed that the total harvest would be the TAC divided by the Faroes share. WGNAS has proposed using a share allocation derived using the same approach and baseline period (1986–1990) as for West Greenland (ICES, 2010a). This gave a potential share allocation of 7.5% to Faroes. Following discussion within NASCO, one Party proposed an alternative baseline period of 1984–1988, which would give a share allocation of 8.4% to Faroes. In the absence of further advice from NASCO, WGNAS has applied a value of 8.4%.

Choice of management units: ICES (2010a) noted that the stock complexes currently used for the provision of NEAC catch advice (southern NEAC and northern NEAC) are significantly larger than each of the six management units used for North American salmon (2SW only) in the catch advice for the West Greenland fishery. Basing an assessment of stock status on these large units greatly increases the risks to individual

NEAC river stocks or groups of stocks that are already in a more depleted state than the average.

For the provision of catch advice on the West Greenland fishery, the total CL for NAC (2SW salmon only) of about 152 000 fish is assessed in six management units, which means that each unit has an average CL of about 25 000 salmon. In contrast, the total CLs for each of the NEAC stock complexes are:

| | |
|--------------------|---------|
| Northern NEAC 1SW– | 158 223 |
| Northern NEAC MSW– | 131 356 |
| Southern NEAC 1SW– | 565 183 |
| Southern NEAC MSW– | 275 549 |

The NEAC stock complexes are therefore between eight and 25 times the size of the average NAC ones. There is also wide variation in the size and status of stocks both within and among the NEAC national stock groups. WGNAS recommended (ICES, 2012a) that the NEAC catch advice should be based on more management units than are used at present, but also noted that there are practical limitations on the extent to which the assessments can be disaggregated, since the availability of information on the composition of the catch at Faroes constrained the selection of management units. In 2013, WGNAS (ICES, 2013) proposed a method to estimate the stock composition of the Faroes catch at a national level based on tag returns and the PFA estimates, but did not consider it appropriate to extend this to stock complexes smaller than this. Genetic stock assignment studies are underway to analyse scale samples collected at Faroes, but these are not expected to facilitate disaggregation below this level. In addition, other parameter values used in the assessment are currently only available for the total fishery and not smaller stock complexes.

In providing indicative catch advice with the new framework, WGNAS considered that it would be informative to managers to provide catch options tables for the ten NEAC countries as well as for the four stock complexes and has therefore run the risk framework using management units based on countries.

Management objectives: The management objectives provide the basis for determining the risks to stocks in each management unit associated with different catch options. The NASCO agreement on the adoption of a Precautionary Approach (NASCO, 1998) calls for the ‘formulation of pre-agreed management actions in the form of procedures to be applied over a range of stock conditions’, indicating that the management objectives (e.g. the required probability of exceeding the CL) should be agreed in advance of specific management proposals being considered.

At the request of NASCO, WGNAS considered the implications of applying probabilities of achieving CLs to separate management units vs. the use of simultaneous probabilities; this issue was outlined in detail in ICES (2013).

The probability of simultaneous attainment of management objectives in a number of separate management units is roughly equal to the product of the probabilities of individual attainment for each management unit. The probability threshold for each individual management unit might reasonably be set at a fixed level unless there are specific reasons for adopting an alternative (e.g. for stock rebuilding). ICES (2012) recommended that an appropriate probability level for individual stock complexes would be 95%. This individual probability level can be applied to each management

unit regardless of the number of units used; however, this is less obvious for the probability of simultaneous attainment, as explained next.

Management decisions for the West Greenland fishery have been based on a 75% probability of simultaneous attainment of CLs. For a given probability of achieving individual stock CLs, the probability of simultaneous attainment decreases rapidly as the number of management units considered increases. For the example of 20 management units (e.g. two age groups from each of ten countries), the use of the simultaneous probability level applied for West Greenland (75%) would correspond to the probability of individual stocks meeting the CLs being 98.6% or higher, assuming the same individual probability for all stocks. The use of a 95% probability level for meeting the CLs individually in the 20 management unit example, implies a simultaneous attainment probability of about 36%, i.e. there would be a 64% chance that at least one stock failed to meet its CL in any given year. On the other hand, the use of a 75% probability of simultaneous attainment could result in a fishery being advised when the individual probability of one management unit is as low as 75% if all the other management units have a 100% chance of meeting the CL (as in that case, the probability of simultaneous attainment would still be 75%). This may not be an acceptable risk for managing multiple river stocks.

WGNAS considered that the probability of simultaneous attainment can provide useful information to managers of the risk of failing to meet CLs in at least one stock in the MSF. However, as the management units being considered by NASCO for managing the MSF at Faroes are still very large and each unit encompasses a large number of individual river stocks, choosing a high probability level (such as 95%) of attaining CLs in individual units would be less risky to individual stocks than the use of a simultaneous attainment objective set at the value used for the West Greenland fishery.

On the basis of these considerations, WGNAS provided both individual probabilities and the probability of simultaneous attainment of the management units in the catch options tables (ICES, 2013). ICES recommends that management decisions should be based principally on a 95% probability of attainment of CLs in each stock complex individually. The simultaneous attainment probability may also be used as a guide, but managers should be aware that this will generally be quite low when large numbers of management units are used (as illustrated above, in the example with 20 management units).

3.4.3.2 Modelling approach for the catch options risk framework

The basic model for assessing each catch option within the risk framework is the same for both stock complexes and at a country level (ICES, 2013). The PFA forecasts derived in the Winbugs model are transferred to the risk framework model run in 'R'. The estimates and distributions of the PFA estimates used in the risk framework are derived by taking the first 50 000 values from the Winbugs posterior forecast simulations. Parameters in the following description that are marked with an '*' in the equations have uncertainty around them generated by means of 50 000 random draws from the annual values observed from the sampling programmes conducted in the Faroes between the 1983/1984 and 1990/1991 fishing seasons. They therefore contribute to the estimation of the probability density function around the potential total harvest arising from each TAC option. When the assessment is run at a national level, the number of draws has to be limited to 25 000 because of memory limitations in 'R'.

The modelling procedure involves:

- estimating the total number of 1SW and MSW salmon that could be killed as a result of any TAC at Faroes, including catches in homewaters;
- adjusting these to their equivalent numbers at the time of recruitment to the Faroes fishery;
- subtracting these from the PFA estimates for maturing and non-maturing 1SW salmon in the appropriate years;
- assessing the results against the SERs (i.e. the CLs adjusted to the time of recruitment to the Faroes fishery).

The TAC option (T) is first divided by the mean weight (Wt^*) of salmon caught in the Faroes fishery to give the number of fish that would be caught, and this value is converted to numbers of wild fish (Nw) by multiplying by one minus the proportion of fish-farm escapees in samples taken from the Faroes catch (pE^*) observed in historical sampling programmes. A correction factor ($C = 0.63$) is applied to the proportion of fish-farm escapees to take account of reductions in the numbers of farm escapees over the past 20 years based on observations in Norwegian coastal waters:

$$Nw = T / Wt^* \times (1 - (pE^* \times C))$$

This value is split into numbers by sea age classes (1SW and MSW) according to the proportion of each age group (pAi^* , where 'i' is 1SW or MSW) observed in historical catch sampling programmes at Faroes. In the past, there has also been a requirement to discard any fish less than 60 cm total length caught in the Faroes fishery, and 80% of these fish were estimated to die, so these mortalities are also added to the 1SW catch. Thus:

$$Nw1SW = Nw_{total} \times pA1SW^* + (Nw_{total} \times pD^* / (1 - pD^*) \times 0.8)$$

and

$$NwMSW = Nw_{total} \times pAMS^*$$

where: pD^* is the proportion of the total catch that is discarded (i.e. fish <60 cm total length).

Further corrections are made to the 1SW and MSW numbers to reduce the 1SW total to take account of the proportion that will not mature as 1SW fish and to add the survivors from this group to the MSW fish in the following year. For the first catch advice year the number added to the MSW total is adjusted to the TAC of the current season (i.e. zero in 2012/2013). Thus:

$$Nw1SW = Nw1SW \times pK^*$$

and

$$NwMSW = NwMSW + Nw1SW \times (1 - pK^*)$$

where ' pK ' is the proportion of 1SW salmon that are expected to mature in the same year (0.78) derived from experimental studies conducted in the 1980s (Youngson and Webb, 1993).

The numbers in each age group are then divided among the management units by multiplying by the appropriate proportions ($pUij$), where 'i' denotes the age groups and 'j' denotes the management units, and each of these values is raised by the Faroes

share allocation (S) to give the total potential harvest (Hij) of fish from each management unit and sea age group:

$$N_{wij} = (N_{wi} \times p_{Uij}) / S$$

Finally, these values are adjusted for natural mortality so that they can be compared with the PFA forecasts and SER values from the mid-date of the fishery to the recruitment date by using an instantaneous monthly rate of mortality of 0.03.

These harvests are then subtracted from the stock forecasts (PFAij) for the management units and sea age groups and compared with the Spawner Escapement Reserves (SERij) to evaluate attainment of the management objective. In practice, the attainment of the management objective is assessed by determining the probability that $PFA_{ij} - H_{ij} - SER_{ij}$ is greater than zero. The SER is the number of fish that need to be alive at the time of the Faroes fishery to meet the CL when the fish return to homewaters; this equals the CL raised by the mortality over the intervening time. CLs and SERs are currently estimated without uncertainty.

3.4.3.3 Input data for the risk framework

The analysis estimates probability of each management unit achieving its SER (the overall abundance objective) for different catch options in the Faroes fishery (from 0 to 200 t). The analysis assumes:

- no fishery operated in the 2012/2013 season;
- the TAC allocated to Faroes is the same in each year and is taken in full;
- homewater fisheries also take their catch allocation in full.

The analysis requires the following input data for the catch that would occur at the Faroes if a TAC was allocated (full details are provided in ICES, 2013):

- mean weights;
- proportion by sea age;
- discard rates (fish less than 60 cm total length);
- proportion of fish-farm escapees;
- composition of catches by management unit;
- proportion of 1SW fish not maturing.

3.4.3.4 Indicative catch advice

Table 3.4.3.4.1 provides an example of catch options for the Faroes fishery for the seasons 2013/2014 to 2015/2016 (ICES, 2013). Equivalent tables were provided for both 1SW and MSW salmon for all NEAC countries, and WGNAS also estimates the exploitation rates that these TAC options would impose on each stock complex or national stock (ICES, 2013). Updated catch options tables for the seasons 2015/2016 to 2017/2018 are provided in Section 3.5 of the 2015 Working Group report (see above).

Table 3.4.3.4.1. Probability of northern and southern NEAC - 1SW and MSW stock complexes achieving their SERs independently and simultaneously for different catch options for the Faroes fishery in the 2013/2014 to 2015/2016 fishing seasons.

| Catch options for 2013/14 season: | TAC option (t) | NEAC-N-1SW | NEAC-N-MSW | NEAC-S-1SW | NEAC-S-MSW | All complexes |
|-----------------------------------|----------------|------------|------------|------------|------------|---------------|
| | 0 | 96.2% | 99.8% | 74.3% | 75.6% | 56.8% |
| | 20 | 96.2% | 99.2% | 74.2% | 69.8% | 52.7% |
| | 40 | 96.2% | 98.2% | 74.2% | 63.9% | 48.2% |
| | 60 | 96.1% | 96.3% | 74.1% | 57.9% | 43.3% |
| | 80 | 96.1% | 93.4% | 74.1% | 52.1% | 38.1% |
| | 100 | 96.1% | 89.3% | 74.0% | 46.6% | 32.9% |
| | 120 | 96.0% | 84.3% | 74.0% | 41.7% | 28.1% |
| | 140 | 96.0% | 78.4% | 73.9% | 36.8% | 23.4% |
| | 160 | 95.9% | 71.6% | 73.9% | 32.5% | 19.2% |
| | 180 | 95.9% | 64.6% | 73.8% | 28.5% | 15.4% |
| | 200 | 95.8% | 57.6% | 73.8% | 25.0% | 12.2% |
| Catch options for 2014/15 season: | TAC option (t) | NEAC-N-1SW | NEAC-N-MSW | NEAC-S-1SW | NEAC-S-MSW | All complexes |
| | 0 | 94.6% | 99.2% | 75.4% | 79.6% | 59.0% |
| | 20 | 94.6% | 98.2% | 75.3% | 75.3% | 55.8% |
| | 40 | 94.6% | 96.6% | 75.3% | 70.8% | 52.0% |
| | 60 | 94.5% | 94.2% | 75.2% | 66.4% | 48.0% |
| | 80 | 94.4% | 90.9% | 75.2% | 61.8% | 43.6% |
| | 100 | 94.4% | 86.8% | 75.1% | 57.3% | 38.9% |
| | 120 | 94.3% | 82.1% | 75.1% | 53.1% | 34.4% |
| | 140 | 94.3% | 76.8% | 75.0% | 49.0% | 30.1% |
| | 160 | 94.3% | 71.2% | 75.0% | 45.0% | 25.9% |
| | 180 | 94.2% | 65.5% | 74.9% | 41.5% | 22.1% |
| | 200 | 94.2% | 59.6% | 74.9% | 38.0% | 18.6% |
| Catch options for 2015/16 season: | TAC option (t) | NEAC-N-1SW | NEAC-N-MSW | NEAC-S-1SW | NEAC-S-MSW | All complexes |
| | 0 | 94.6% | 98.5% | 70.1% | 79.7% | 55.2% |
| | 20 | 94.6% | 97.2% | 70.1% | 76.0% | 52.4% |
| | 40 | 94.5% | 95.1% | 70.0% | 72.2% | 49.2% |
| | 60 | 94.5% | 92.3% | 70.0% | 68.4% | 45.6% |
| | 80 | 94.5% | 89.0% | 69.9% | 64.6% | 41.9% |
| | 100 | 94.4% | 85.0% | 69.9% | 60.7% | 38.0% |
| | 120 | 94.4% | 80.6% | 69.8% | 57.1% | 34.2% |
| | 140 | 94.3% | 75.7% | 69.8% | 53.5% | 30.4% |
| | 160 | 94.3% | 70.6% | 69.7% | 50.0% | 26.7% |
| | 180 | 94.2% | 65.4% | 69.7% | 46.8% | 23.4% |
| | 200 | 94.2% | 60.4% | 69.7% | 43.7% | 20.4% |

3.5 Development of indicator frameworks to identify significant changes in previously provided multiannual management advice

3.5.1 Background

In support of the multiannual management advice that is provided for all three NASCO Commission Areas, NASCO asked ICES to provide an assessment of the minimal information needed to signal an unforeseen change in productivity for stocks

contributing to fisheries within each Commission area. A particular concern was that an increase in productivity may alter the reliability of the previously provided multiyear catch options and could result in unrealised harvest within various mixed-stock fisheries. Initial progress on this issue was presented to WGNAS in 2006 (ICES, 2006) and further developments were made by the Study Group on Establishing a Framework of Indicators of Salmon Stock Abundance [SGEFISSA] which met in November 2006 (ICES, 2007b) and reported to WGNAS in 2007 (ICES, 2007a). This resulted in the development of a suggested framework (Framework of Indicators - FWI) which could be used to indicate if any significant change in the status of stocks had occurred and thus confirm whether the previously provided multi-annual management advice was still appropriate.

The initial FWI was developed with both the Greenland and Faroes fisheries in mind, although the methodology only proved suitable for the West Greenland fishery and an alternative approach was subsequently developed for the NEAC area (ICES, 2011; 2012a; 2013). Thus, FWIs are now routinely applied in the interim (non-assessment) years of multiyear agreements for both NAC and NEAC to facilitate the management of the West Greenland and Faroes fisheries respectively. Both operate according to the timeline outlined in Figure 3.5.1.1. Outline descriptions of the two different schemes are provided below.

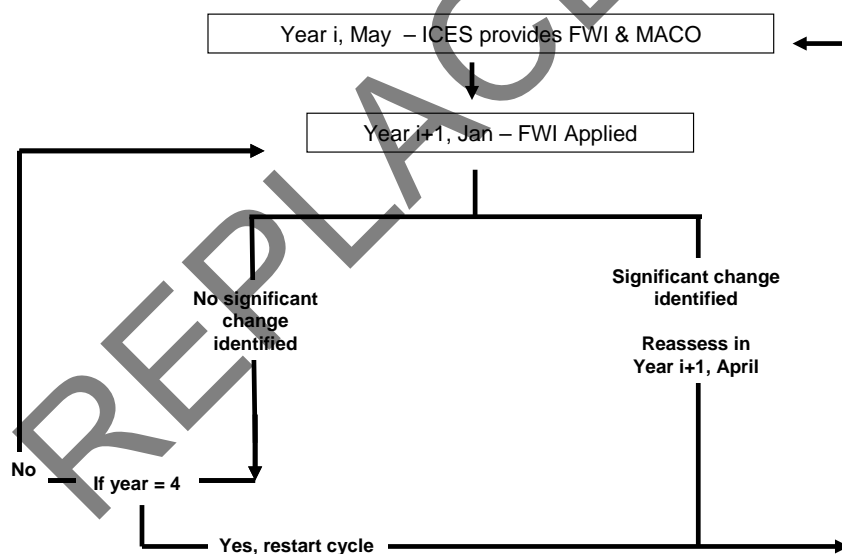


Figure 3.5.1.1. Timeline for employment of the Framework of Indicators (FWI). In Year i , ICES provides an updated FWI which re-evaluates the updated datasets and is summarized in an Excel worksheet. In January of Year $i+1$ the FWI is applied and two options are available depending on the results. If no significant change is detected, no re-assessment is necessary and the cycle continues to Year $i+2$. If no significant change is detected in Year $i+2$, the cycle continues to Year $i+3$. If a significant change is detected in any year, then reassessment is recommended. In that case, ICES would provide an updated FWI the following May. ICES would also provide an updated FWI if year equals 4. [MACO = multi-annual catch options].

3.5.2 Framework of Indicators (FWI) for the West Greenland Fishery

The process for developing and applying the FWI for the Greenland fishery consists of six general steps:

- Definition of a significant change - Define measurable criteria for what the statement “a significant change in the previously provided multi-annual management advice” represents.
- Evaluating historical relationships between indicators and variables of interest - Define and evaluate the historical relationships between numerous indicators and the variable of interest for individual rivers across all stock complexes.
- Establishing threshold values - Define the threshold level (i.e. variable of interest level) that will satisfy the management objectives for each stock complex.
- Decision rule determinations - Define and apply a standardised approach for determining the appropriate decision rule value. The decision rule should provide a signal if the variable of interest will be greater than or less than the threshold level with high precision.
- Combining Indicators within the Framework - Define and apply a standardised approach for combining indicator datasets within and across stock complexes for future comparison against contemporary indicator values.
- Applying the FWI - Define and apply a standard approach to input contemporary indicator values into the FWI to determine if there is likely to be a significant change in the previously provided management advice.

Each of these is considered in turn; full details are available in ICES (2007b).

3.5.2.1 Definition of a significant change

A significant change in the previously provided multiannual management advice is regarded as an unforeseen change in stock status that would alter the previously provided advice based on analysis of current population data obtained from various monitored populations across the North Atlantic. This would be indicated by a situation where stock abundance has increased to a level where a fishery could be recommended when no catch had previously been advised, or a decrease in stock abundance when catch options had been chosen.

For the fishery at West Greenland, ICES would recommend that a harvestable surplus exists within the West Greenland stock complex if there was a high probability (75%) that the following three objectives could be met simultaneously:

The conservation limits of the four northern regions of North America (Labrador, Newfoundland, Québec, and Gulf) were achieved.

There was a 25% increase in returns to the Scotia-Fundy region relative to the mean returns for the 1992–1996 period. For the USA, the management objective was revised in 2014 to correspond to recover objectives defined in the recovery plan for endangered Atlantic salmon stocks in the USA (ICES, 2014), this now requires that estimated 2SW adult returns are 4549 or greater.

The conservation limit for the Southern NEAC MSW complex was achieved.

3.5.2.2 Evaluating historical relationships between indicators and variable of interest

A number of variables were considered for inclusion as indicators in the FWI, but only two were considered sufficiently informative to be carried forward into the framework: adult returns (returns, catch or estimated PFA) and return rates (i.e. smolt survival

rates, marine survival). These are available, by sea age class, for a number of monitored rivers throughout the North Atlantic and can be directly related to the management objectives for the fishery.

3.5.2.3 Establishing threshold values

In keeping with the 75% probability of meeting or exceeding the objectives for the West Greenland catch options (see above), the 25th percentile of the return estimates of the six areas in North America are compared to the corresponding 2SW conservation limits of the four northern areas of North America, to the 25% increase objective for the Scotia-Fundy area, and to management objective of achieving 4549 or greater 2SW adult returns for the USA. For the southern NEAC non-maturing component, the 25th percentile of the PFA estimate of the southern NEAC non-maturing complex is compared to the spawning escapement reserve (SER) for the southern NEAC non-maturing complex.

3.5.2.4 Decision rule determinations

The procedure for analysing the relationships between the indicators and the returns of 2SW salmon or the PFA estimates was originally suggested by ICES (2006). The individual river catches, returns or return rates are lagged to correspond to the same smolt cohort for the 2SW returns to North America or to the PFA estimates for NEAC complexes. Bivariate plots of each indicator dataset relative to the 2SW returns or the PFA estimates are prepared. Upper and lower halves are defined by the management objective value for the corresponding geographic area in North America or the NEAC stock complexes as outlined above. Estimates of returns of 2SW or PFA estimates in the upper half correspond to years when the returns or PFA exceed the management objectives. Points in the lower half correspond to years when the returns or PFA are less than the management objective.

Left and right halves are defined by a sliding rule along the indicator range. An objective function that maximises the number of correct assignments (true highs and true lows) is used to define the indicator decision rule. The objective function also minimises the number of incorrect assignments (false highs and false lows, Figure 3.5.2.4.1).

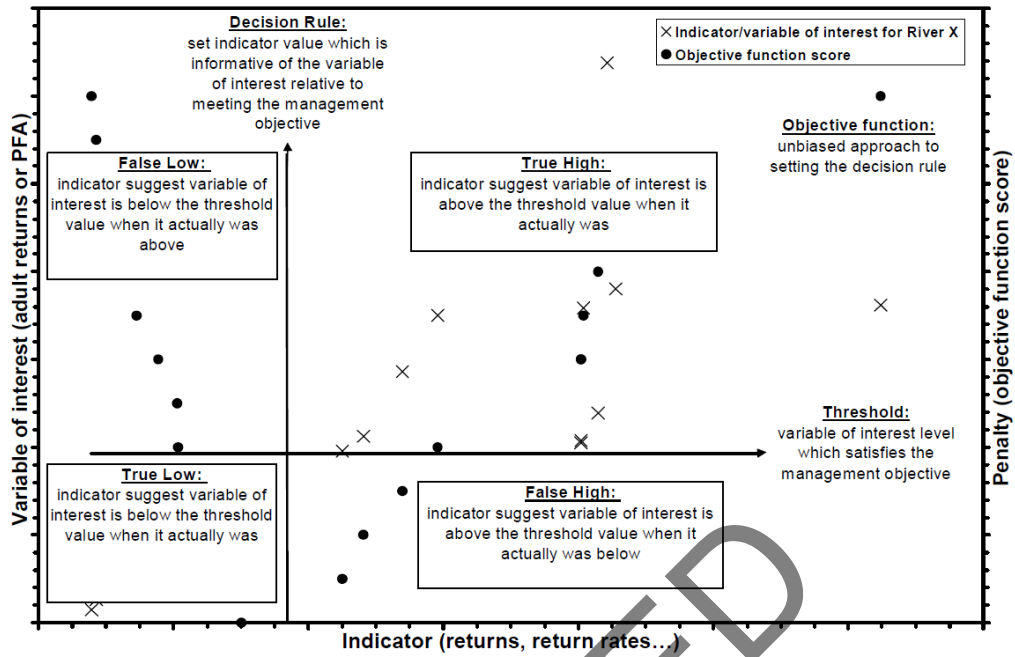


Figure 3.5.2.4.1. Example of Indicator/Variable of Interest exploratory graph identifying the threshold value, decision rule, penalty function and the four states (true high, true low, false high and false low).

The value of the indicator variable that minimises the sum of the penalty scores (i.e. minimises the number of incorrect assignments) is assigned as the decision rule for that dataset. Equal penalty weights are assigned to false highs (lower right quadrant) and false lows (upper left quadrant). Correct assignments are scored as zero. In the case when multiple minima occurred, the lowest indicator value among the low minima values is chosen.

Indicators are retained in the framework when they are evaluated as being informative of the magnitude of returns or PFA relative to the management objectives. These informative indicator datasets also have to meet the following two criteria to be retained:

- Expectation that the indicator variable would be available in future (in January), and
- A minimum of five observations are present in each of the correct quadrats (true low; true high).

3.5.2.5 Combining Indicators within the Framework

The probabilities of correct assignments are calculated for each of the true low and true high states for each of the indicator datasets retained. The respective probabilities correspond to the ratio of the correct assignment to all observations within the respective low and high indicator halves:

$$P(\text{State}_{\text{low}} \mid \text{Indicator}_{\text{low}}) \text{ (i.e. true low)} = N(\text{State}_{\text{low}} \mid \text{Indicator}_{\text{low}}) / N \text{ Indicator}_{\text{low}}$$

$$P(\text{State}_{\text{high}} \mid \text{Indicator}_{\text{high}}) \text{ (i.e. true high)} = N(\text{State}_{\text{high}} \mid \text{Indicator}_{\text{high}}) / N \text{ Indicator}_{\text{high}}$$

Indicator datasets are then pooled according to management objective/stock complex groupings. Each NAC stock complex (n=6) and the Southern NEC non-maturing stock complex are pooled separately as these stock complexes relate to the management objectives for the West Greenland fishery.

3.5.2.6 Applying the FWI

To apply the FWI, the most recent year's indicator value for each of the retained indicator datasets is compared to the decision rule as determined from the historical datasets. If the contemporary indicator value is low relative to the decision rule, it is assigned a value of -1. If the value is high, it is assigned a score of +1. Multiple indicators within the stock complex groupings are then combined by arithmetic average of the product of the indicator value (-1, +1) and the probability of a correct assignment corresponding to the true low or true high states. An average geographic area or stock complex score equal to or greater than zero suggests there is a likelihood of meeting the management objective for that grouping based on the historic relation between the variable of interest (adults returns or PFA) and the indicators evaluated.

If the scores for all the groupings within a fishery complex are greater than zero, then there is a likelihood that all the management objectives for that fishery will be met. Under that scenario, the multiyear management advice should be reassessed. When the score(s) for one or more of the groupings is less than zero, there is unlikely to be a significant change in the management advice and there would be no need for a reassessment.

SGEFISSA (ICES, 2007b) developed a spreadsheet template FWI (see example at Figure 3.5.2.6.1) in which the underlying variable of interest/ indicator dataset relationships and decision rules are summarised and collated according to the specific management objectives for each fishery. This provides one of two conclusions for the user:

- 1) No significant change identified by the indicators;
- 2) Reassess.

| Catch Advice | | Catch option > 0 (Yes = 1, No = 0) | | | 0 | | | | | | |
|------------------------------------------------|-------------------------------------|---------------------------------------|--------------------------|-----------|----------|-----------|-----------------|-----------------------------------|-----------------|---------------------------|--|
| Overall Recommendation | | | | | | | | | | | |
| No Significant Change Identified by Indicators | | | | | | | | | | | |
| Geographic Area | River/ Indicator | 2011 Value | Ratio Value to Threshold | Threshold | True Low | True High | Indicator State | Probability of Correct Assignment | Indicator Score | Management Objective Met? | |
| USA | Penobscot 2SW Returns | 2368 | 167% | 1415 | 100% | 92% | 1 | 0.92 | 0.92 | | |
| | Penobscot 1SW Returns | 741 | 197% | 377 | 83% | 88% | 1 | 0.88 | 0.88 | | |
| | Penobscot 2SW Survival (%) | 0.39 | 170% | 0.23 | 100% | 60% | 1 | 0.6 | 0.6 | | |
| | Penobscot 1SW Survival (%) | 0.12 | 133% | 0.09 | 85% | 73% | 1 | 0.73 | 0.73 | | |
| | Narraguagus Returns | 196 | 196% | 100 | 95% | 61% | 1 | 0.61 | 0.61 | | |
| | possible range | | | | -0.93 | 0.75 | | | | | |
| | Average | | 173% | | | | | | 0.75 | Yes | |
| Scotia-Fundy | Saint John Return Large | 294 | 9% | 3 329 | 96% | 100% | -1 | 0.96 | -0.96 | | |
| | Lahave Return Large | 146 | 51% | 285 | 77% | 85% | -1 | 0.77 | -0.77 | | |
| | St. Mary's Return Large | 14 | 6% | 221 | 100% | 73% | -1 | 1 | -1 | | |
| | North Return Large | 1 193 | 168% | 712 | 95% | 67% | 1 | 0.67 | 0.67 | | |
| | Saint John Return 1SW | 582 | 26% | 2 276 | 86% | 80% | -1 | 0.86 | -0.86 | | |
| | LaHave Return 1SW | 565 | 34% | 1 679 | 94% | 67% | -1 | 0.94 | -0.94 | | |
| | St. Mary's Return 1SW | 331 | 16% | 2 038 | 95% | 93% | -1 | 0.95 | -0.95 | | |
| | Saint John Survival 2SW (%) | 0.13 | 59% | 0.22 | 95% | 81% | -1 | 0.95 | -0.95 | | |
| | Lahave Survival 2SW (%) | 0.88 | 367% | 0.24 | 81% | 81% | 1 | 0.81 | 0.81 | | |
| | Saint John Survival 1SW (%) | 0.12 | 16% | 0.76 | 86% | 73% | -1 | 0.86 | -0.86 | | |
| | Lahave Survival 1SW (%) | 0.72 | 50% | 1.44 | 92% | 78% | -1 | 0.92 | -0.92 | | |
| | Liscomb Survival 2SW (%) | 0.03 | 60% | 0.05 | 86% | 91% | -1 | 0.86 | -0.86 | | |
| | East Sheet Harbour Survival 2SW (%) | 0.005 | 25% | 0.02 | 67% | 82% | -1 | 0.67 | -0.67 | | |
| | possible range | | | | -0.88 | 0.81 | | | | | |
| | Average | | 68% | | | | | | -0.64 | No | |
| Gulf | Miramichi Return 2SW | 28 977 | 183% | 15 800 | 100% | 85% | 1 | 0.85 | 0.85 | | |
| | Miramichi Return 1SW | 45 880 | 110% | 41 790 | 89% | 67% | 1 | 0.67 | 0.67 | | |
| | possible range | | | | -0.95 | 0.76 | | | | | |
| Average | | 147% | | | | | | 0.76 | Yes | | |
| Quebec | Cascapédia Return Large | 3 815 | 167% | 2 280 | 69% | 92% | 1 | 0.92 | 0.92 | | |
| | Bonaventure Return Large | 1 259 | 85% | 1 479 | 75% | 81% | -1 | 0.75 | -0.75 | | |
| | Grande Rivière Return Large | 533 | 121% | 442 | 100% | 94% | 1 | 0.94 | 0.94 | | |
| | Saint-Jean Return Large | 688 | 91% | 758 | 86% | 89% | -1 | 0.86 | -0.86 | | |
| | Dartmouth Return Large | 1 171 | 155% | 756 | 86% | 89% | 1 | 0.89 | 0.89 | | |
| | Madeleine Return Large | 996 | 153% | 653 | 70% | 93% | 1 | 0.93 | 0.93 | | |
| | Sainte-Anne Return Large | 871 | 201% | 433 | 67% | 88% | 1 | 0.88 | 0.88 | | |
| | Godbout Return Large | 694 | 108% | 641 | 86% | 100% | 1 | 1 | 1 | | |
| | De la Trinité Return Large | 317 | 82% | 385 | 75% | 100% | -1 | 0.75 | -0.75 | | |
| | York Return Return Large | 1 585 | 113% | 1405 | 63% | 83% | 1 | 0.83 | 0.83 | | |
| | Grande Rivière Return Small | 237 | 119% | 199 | 59% | 80% | 1 | 0.8 | 0.8 | | |
| | Saint-Jean Return Small | 343 | 87% | 394 | 53% | 80% | -1 | 0.53 | -0.53 | | |
| | Godbout Return Small | 623 | 123% | 508 | 85% | 92% | 1 | 0.92 | 0.92 | | |
| | De la Trinité Return Small | 949 | 238% | 399 | 89% | 83% | 1 | 0.83 | 0.83 | | |
| | De la Trinité Survival Large (%) | 0.76 | 155% | 0.49 | 88% | 96% | 1 | 0.96 | 0.96 | | |
| | De la Trinité Survival Small (%) | 2.54 | 170% | 1.49 | 63% | 89% | 1 | 0.89 | 0.89 | | |
| | Saint-Jean Survival Small (%) | 1.86 | 258% | 0.72 | 100% | 64% | 1 | 0.64 | 0.64 | | |
| | possible range | | | | -0.77 | 0.88 | | | | | |
| | Average | | 143% | | | | | | 0.50 | Yes | |
| Newfoundland | Exploits Return Small | 34 085 | 137% | 24 924 | 83% | 56% | 1 | 0.56 | 0.56 | | |
| | Middle Brook Return Small | 2 642 | 141% | 1 868 | 84% | 63% | 1 | 0.63 | 0.63 | | |
| | Torrent Return Small | 2 784 | 67% | 4 154 | 94% | 64% | -1 | 0.94 | -0.94 | | |
| | possible range | | | | -0.87 | 0.61 | | | | | |
| Average | | 115% | | | | | | 0.08 | Yes | | |
| Labrador | possible range | | | | | | | | | | |
| | Average | | | | | | | | NA | Unknown | |
| Southern NEAC | possible range | | | | | | | | | | |
| | Average | | | | | | | | NA | Unknown | |

Figure 3.5.2.6.1. Framework of indicators spreadsheet for the West Greenland fishery. For illustrative purposes, the 2011 value of returns or survival rates for the 40 retained indicators is entered in the cells corresponding to the annual indicator variable values.

If no significant change has been identified by the indicators, then the multiyear catch advice for the year of interest could be retained. If a significant change is signalled by the indicators, the response is to reassess.

The framework spreadsheet is designed to capture both fishing and non-fishing scenarios:

- Multiyear advice provides no catch options greater than zero but indicators are suggesting that the management objectives may be met (conclusion: Reassess);
- Multiyear advice provides catch options greater than zero but the indicators suggest the management objectives may not be met (conclusion: Reassess).

There are two steps required by the user to run the framework. The first step in the framework evaluation is to enter the catch advice option (i.e. tonnes of catch) for the

West Greenland fishery. This feature provides the two way evaluation of whether a change in management advice may be expected and a reassessment would be required. The second step is to enter the values for the indicator variables in the framework for the year of interest. The spreadsheet evaluation update is automated and the conclusion is shown in the row underneath “Overall Recommendation”.

The conclusions from the framework evaluation are based on whether there is simultaneous achievement of the management objectives in the six stock areas of North America and the southern NEAC 1SW non-maturing complex. If there are no indicator variables for a geographic area, the attainment of the management objectives is evaluated as unknown and that area or complex is not used in the decision structure of the framework.

Within the geographic areas for which indicator variables are retained, all the available indicators are used to assess the indicator score. If an update value for an indicator variable is not available for the year of interest, the indicator variable is not used to quantify the indicator score for that area.

The West Greenland FWI was updated during the 2015 WGNAS meeting and an updated spreadsheet produced. Details are provided in Section 5.8 of the Working Group report (see above).

3.5.3 Framework of Indicators (FWI) for the Faroes Fishery

3.5.3.1 Background

The original FWI applied to the West Greenland fishery (ICES, 2007b) was not applicable for the Faroese fishery for a number of different reasons. Among these were the lack of quantitative catch advice, the absence of specific management objectives and a sharing agreement for this fishery and the fact that none of the available indicator datasets met the criteria for inclusion in the FWI.

In 2011, WGNAS re-evaluated the approach for developing a FWI for the Faroese fishery (ICES, 2011). Since over the available time-series the PFA estimates for the NEAC stock complexes have predominately remained above the SER, the Working Group suggested a different set of decision rules for this FWI. It was suggested that the status of stocks should be re-evaluated if the FWI suggests that the PFA estimates are deviating substantially from the median values from the forecast.

3.5.3.2 Description of the FWI

It was initially suggested that the 95% confidence interval range for the mean of the indicator prediction, relative to the median forecast value, be used to compute the decision thresholds for whether the indicator suggests a reassessment or not (ICES, 2011). The limits should be computed at the median values of the PFA forecasts in each of the years in multiyear advice. However, the 95% criterion was subsequently re-examined (ICES, 2012a) and it was recommended that the upper and lower 75% confidence limits of the individual predictions be used for comparison (Figure 3.5.3.2.1). WGNAS recognised that this was a relaxation of the decision rule suggested in 2011, and will lead to a larger interval, and thus a lower chance of a reassessment than the approach suggested in 2011. However, this was considered to be a more realistic confidence level given the relatively wide variability of the indicator datasets, and was also consistent with the approach adopted by NAC.

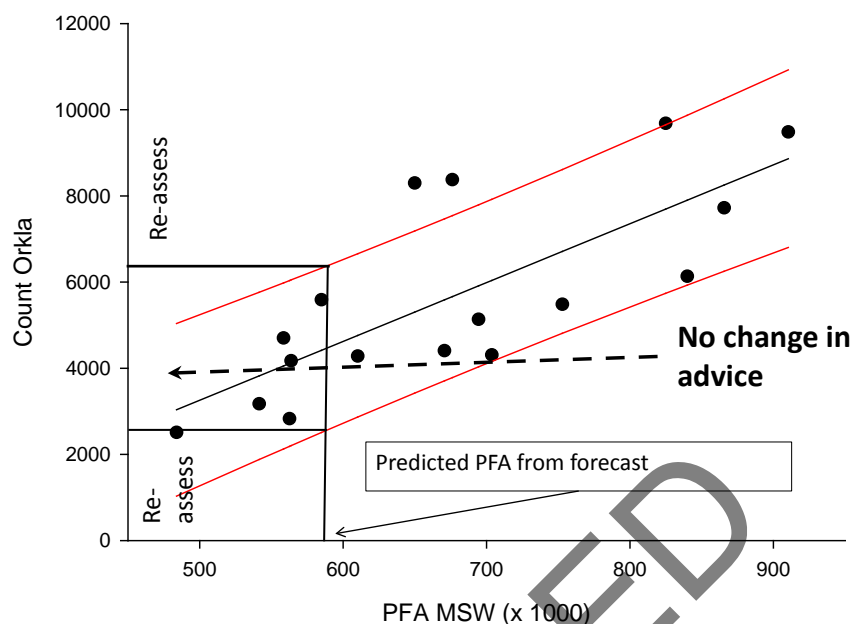


Figure 3.5.3.2.1. Example of how the reassessment intervals for the indicators are computed. The values of an indicator (counts) are plotted against the PFA. The regression line is shown in black and 75% confidence limits for the individual estimates are shown in red. From the forecasted PFA in the year in question the values of the indicator corresponding to the upper and lower 75% confidence interval are estimated. If the indicator value falls outside these limits a reassessment is recommended by this particular indicator.

When the stocks are divided into smaller management units, potential indicators for each management unit become relatively scarce. Therefore, the Working Group recommended that the FWI be regressed against the stock complexes that they belong to. For example, MSW indicators from Norway should be regressed against PFA MSW for northern NEAC.

In 2012, the FWI was applied as a two-tailed test (ICES, 2012a). However, it was subsequently agreed that, in the event of a closed fishery, the indicators should only be compared to the upper 75% confidence limit (i.e. a one-tailed test). This means that for a closed fishery, a reassessment is only triggered where the forecast appeared to be an underestimate and there may be a possibility of a harvest being denied. In the case of an open fishery they should be compared to both the upper and lower 75% confidence limits. In this case, if the FWIs suggest that the forecasted PFA is either an underestimation or an overestimation of the realised PFA in any of the four stock complexes, then this should trigger a reassessment.

WGNAS developed a FWI spreadsheet (ICES, 2011) to provide an automatic evaluation of the need for a reassessment once the new indicator values are available in January; this has been updated in subsequent years (ICES, 2012a; 2013). An example spreadsheet is provided at Figure 3.5.3.2.2.

The following summarizes the main steps performed by the spreadsheet following updating of the relevant data for the variable of interest by adding the latest year's number:

- Regression analysis with the dataset x to determine its power to predict PFA in the forecasted years.
- Calculation of the 75% confidence intervals of individual predictions of the regression for dataset x . An indicator value below the 75% individual confidence interval (CI) is interpreted as indicative of an overestimation of the PFA, while a point above the 75% individual confidence interval is interpreted as indicative of an underestimation of PFA.
- A dataset is considered informative and should be kept as an indicator in the FWI if the following conditions are met: sample size (n) ≥ 10 ; $r^2 \geq 0.2$; dataset updated annually and new value available by January 15. Datasets that do not meet these criteria are discarded.
- Apply a binary score to each indicator value. Thus, for dataset x , if the current year's indicator value is outside the 75% individual regression point estimate CI (below or above) then that indicator receives a score of 1. If the indicator is within the 75% CI, then the indicator receives a score of -1. In the absence of an indicator data point for any year, a score of zero is applied. Whether the indicator value is above or below the upper and lower CI values is checked separately in two spreadsheet columns and a decision whether the indicator value is within the CI is assessed by combining the information in the two columns.
- Separate columns are used to sum the scores for all the indicator datasets within each stock complex. This is done separately for points that fall above the CI and those that fall below. In the case of a two-sided approach (open fishery), if the sum of these columns is ≥ 0 , then the spreadsheet signals "REASSESS"; if the sum is < 0 , then it signals "No significant OVERestimation of PFA identified by indicators, do not reassess" for indicator values that fall below the CI, and "No significant UNDERestimation of PFA identified by indicators, do not reassess" for indicator values that are above the CI. In case of a one-sided approach (closed fishery), only underestimation will signal a "REASSESS".
- FWI results are generated for each stock complex (northern NEAC maturing and non-maturing, and southern NEAC maturing and non-maturing). A score of ≥ 0 for any of these stock complexes would signal a reassessment.

WGNAS reassessed the effects of applying stricter criteria than $r^2 \geq 0.2$ for inclusion of indicators in the FWI. As stricter criteria are used, the number of indicators included reduces rapidly. It was therefore concluded to keep the criterion of $r^2 \geq 0.2$ in order to obtain a sufficient number of indicators to be able to use the FWI even in the event of one or more indicators being unavailable by the time the FWI is applied each year. The r^2 value of 0.2 corresponds to a value slightly lower than what is considered to be a "large" effect size ($r = 0.5$, $r^2 = 0.25$) by Cohen (1988). Although a criterion of $r^2 \geq 0.2$ gives each indicator little predictive power alone (Prairie, 1996), the approach of using a suite of indicators is more similar to meta-analysis (Rosenthal, 1984) meaning that the outcome of the FWI is not dependent on the result of one indicator in isolation, but rather on the combined performance of the indicator set.

The Faroes FWI was updated during the 2015 WGNAS meeting and an updated spreadsheet produced. Details are provided in Section 3.7 of the Working Group report (see above). In evaluating all the time-series, it was noted that the lower 12.5 % CL, which is used to determine which indicator values are outside the 75% CI on the lower

side, was negative for some regression relationships for predicted PFA values in 2015 and 2016. Since this would invalidate the use of such indicators (they would not indicate that predicted PFA values are overestimates regardless of how small they are), an additional (fourth) criterion was established as a requirement for including time-series in the FWI. This requires that the lower 12.5% confidence limit for an indicator time-series should be positive for any values of PFA included in the FWI.

| FWI NEAC | | 2012 | | Indicators suggest: | | REASSESS | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
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| An example | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | </ |

Figure 3.5.3.2.2. Output of the spreadsheet for the test of FWIs for NEAC for 2012 based on the values of the indicators from 2011. Because the indicators suggest that the forecast for northern NEAC MSW was an underestimate, the overall advice from the spreadsheet is reassess.

4 References

- Amiro, P., Hubley, B., Gibson, J. and Jones, R. 2008. Estimates of returns and escapements of Atlantic salmon to Salmon Fishing Areas 19, 20 and 21, Nova Scotia, and update of returns and escapements to SFA 23, southern New Brunswick, 1970 to 2007. ICES North Atlantic Salmon Working Group, Working Paper 2008/014.
- Anderson, T.C. 1985. The Rivers of Labrador. Can. Spec. Publ. Fish. Aquat. Sci. 81: 389p.
- Aprahamian, M.W., Davidson, I.C. and Cove, R.J. 2008. Life history changes in Atlantic salmon from the river Dee, Wales. *Hydrobiologia* 602: 61–78.

- Ash, E.G.M. and O'Connell, M.F. 1987. Atlantic salmon fishery in Newfoundland and Labrador, commercial and recreational, 1985. Can. Data Rep. Fish. Aquat. Sci. 672: v + 284 p.
- Caron, F., Fontaine, P.-M. and Picard, S.-E. 1999. Seuil de conservation et cible de gestion pour les rivières à saumon (*Salmon salar*) du Québec. Faune et parcs Québec, Direction de la faune et des habitats. ISBN: 2-550-35636-5, 48 pp.
- Cefas and Environment Agency. 2013. Annual assessment of salmon stocks and fisheries in England and Wales, 2012. ISBN: 978 0 907545 74 3, 138pp.
- Chaput, G., Legault, C.M., Reddin, D.G., Caron, F. and Amiro, P.G. 2005. Provision of catch advice taking account of non-stationarity in productivity of Atlantic salmon (*Salmo salar* L.) in the Northwest Atlantic. *ICES Journal of Marine Science* 62: 131–143.
- Chaput, G. 2012. Overview of the status of Atlantic salmon (*Salmo salar* L.) in the North Atlantic and trends in marine mortality. *ICES Journal of Marine Science* 69: 1538–1548.
- Cohen, J. 1988. Statistical power analysis for the behavioral sciences. Lawrence Erlbaum Associates, Publishers, Hillsdale, New Jersey.
- Crozier, W.W., Potter, E.C.E., Prevost, E., Schon, P.J., and O'Maoiléidigh, N. 2003. A co-ordinated approach towards development of a scientific basis for management of wild Atlantic salmon in the North-East Atlantic (SALMODEL). An EU Concerted Action - Quality of Life and Management of Living Resources Key Action 5: Sustainable agriculture, fisheries and forestry, and integrated development of rural areas including mountain areas. Contract No. QLK5-CT1999-01546, 431 pp.
- Davidson, I.C. and Hazelwood, M.S. 2005. Effects of climate change on salmon fisheries. Science Report W2-047/SR, Environment Agency, Bristol, 52 pp.
- Decisioneering. 1996. Crystal Ball-Forecasting and risk analysis for spreadsheet users (Version 4.0). 286 pp.
- Doubleday, W.G., Rivard, D.R., Ritter, J.A., and Vickers, K.U. 1979. Natural mortality rate estimates for North Atlantic salmon in the sea. International Council for the Exploration of the Sea CM 1979/M: 26, 15 pp.
- Fairchild, W.L., Brown, S.B. and Moore, A. 2002. Effects of freshwater contaminants on marine survival in Atlantic salmon. NPAFC: 30–32.
- Friedland, K.D. 1998. Ocean climate influences on critical Atlantic salmon (*Salmo salar*) life history events. *Canadian Journal of Fisheries and Aquatic Sciences* 55 (Suppl. 1): 119–130.
- Friedland, K.D., Reddin, D.G. and Castonguay, M. 2003. Ocean thermal conditions in the post-smolt nursery of North American Atlantic salmon. *ICES Journal of Marine Science* 60: 343–355.
- Friedland, K.D., MacLean, J.C., Hansen, L.P., Peyronnet, A.J., Karlsson, L., Reddin, D.G., Ó Maoiléidigh, N. and McCarthy, J.L. 2009. The recruitment of Atlantic salmon in Europe. *ICES Journal of Marine Science* 66: 289–304.
- Friedland, K.D., Shank, B.V., Todd, C.D., McGinnity, P. and Nye, J.A. 2013. Differential response of continental stock complexes of Atlantic salmon (*Salmo salar*) to the Atlantic Multidecadal Oscillation. *Journal of Marine Systems* (2013), <http://dx.doi.org/10.1016/j.jmarsys.2013.03.003>.
- Glover, K.A., Quintela, M., Wennevik, V., Besnier, F., Sorvik, A.G.E. and Skaala, O. 2012. Three Decades of Farmed Escapees in the Wild: A spatio-temporal analysis of Atlantic salmon population genetic structure throughout Norway. *Plos One* 7(8): e43129. doi:10.1371/journal.pone.0043129.
- Glover, K.A., Pertoldi, C., Besnier, F., Wennevik, V., Kent, M. and Skaala, O. 2013. Atlantic salmon populations invaded by farmed escapees: quantifying genetic introgression with a Bayesian approach and SNPs. *BMC Genetics* 14: 74. doi: 10.1186/1471-2156-14-74.

- Hansen, L.P., Døving, K.B. and Jonsson, B. 1987. Migration of farmed adult Atlantic salmon with and without olfactory sense, released on the Norwegian coast. *Journal of Fish Biology* 30: 713–721.
- Hansen, L.P., Jacobsen, J.A. and Lund, R.A. 1997. The incidence of reared Atlantic salmon (*Salmo salar* L.) of fish farm origin at West Greenland. *ICES Journal of Marine Science* 54: 152–155.
- Hansen, L.P. and Quinn, T.P. 1998. The marine phase of Atlantic salmon (*Salmo salar*) life cycle, with comparison to Pacific salmon. *Canadian Journal of Fisheries and Aquatic Sciences* 55 (Suppl. 1): 104–118.
- Hansen, L.P., Jacobsen, J.A. and Lund, R.A. 1999. The incidence of escaped farmed Atlantic salmon, *Salmo salar*, in the Faroese fishery and estimates of catches of wild salmon. *ICES Journal of Marine Science* 56: 200–206.
- Hansen, L.P. and Jacobsen, J.A. 2003. Origin, migration and growth of wild and escaped farmed Atlantic salmon, *Salmo salar* L., in oceanic areas north of the Faroe Islands. *ICES Journal of Marine Science* 60: 110–119.
- Hutchings, J.A. and Jones, M.E.B. 1998. Life history variation and growth rate thresholds for maturity in Atlantic salmon, *Salmo salar*. *Canadian Journal of Fisheries and Aquatic Sciences* 55 (Suppl. 1): 22–47.
- ICES. 1984. Report of the Working Group on North Atlantic Salmon. ICES CM 1984/Assess: 16, 54 pp.
- ICES. 1991. Report of the Working Group on North Atlantic Salmon. Copenhagen, 14–21 March 1991. ICES, CM 1991/Assess: 12, 157 pp.
- ICES. 1993a. Report of the North Atlantic Salmon Working Group. Copenhagen, 5–12 March 1993. ICES, CM 1993/Assess: 10.
- ICES. 1993b. Report of the Study Group on North American Salmon Fisheries. Woods Hole, Mass., 15–19 February 1993. ICES, CM 1993/Assess: 9.
- ICES. 1994. Report of the Working Group on North Atlantic Salmon. Reykjavik, 6–15 April 1994. ICES, Doc. CM 1994/Assess: 16, Ref. M, 182 pp.
- ICES. 1995. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 3–12 April 1995. ICES, Doc. CM 1995/Assess: 14, Ref. M, 191 pp.
- ICES. 1996. Report of the Working Group on North Atlantic salmon, Moncton, Canada, 10–19 April 1996. ICES CM 1996/Assess: 11, Ref: M, 228 pp.
- ICES. 1998. Report of the Working Group on North Atlantic Salmon. Copenhagen, 14–23 April 1998. ICES, CM 1998/ACFM: 15.
- ICES. 1999. Report of the Working Group on North Atlantic Salmon. Copenhagen, 12–22 April 1999. ICES, CM 1999/ACFM: 14.
- ICES. 2000. Report of the Working Group on the North Atlantic Salmon. ICES Headquarters, Copenhagen, April 3–13, ICES CM 2000/ACFM: 13, 301 pp.
- ICES. 2001. Report of the Working Group on North Atlantic Salmon. Aberdeen, 2–11 April 2001. ICES CM 2001/ACFM: 15, 290 pp.
- ICES. 2002. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 3–13 April 2002. ICES CM 2002/ACFM: 14, 299 pp.
- ICES. 2003. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 31 March–10 April 2003. ICES CM 2003/ACFM: 19, 297 pp.
- ICES. 2005. Report of the Working Group on North Atlantic Salmon. Nuuk, Greenland 4–14 April. ICES CM 2005/ACFM: 17, 290 pp.
- ICES. 2006. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 4–13 April. ICES CM 2006/ACFM: 23, 254 pp.

- ICES. 2007a. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 11–20 April. ICES CM 2007/ACFM: 13, 253 pp.
- ICES. 2007b. Study Group on Establishing a Framework of Indicators of Salmon Stock Abundance (SGEFISSA). ICES CM 2007/DFC: 01, 71 pp.
- ICES. 2008. Report of the Working Group on North Atlantic Salmon. Galway, Ireland 1–10 April. ICES CM 2008/ACOM: 18, 235 pp.
- ICES. 2009a. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 30 March–8 April 2009. ICES CM 2009/ACFM: 06, 283 pp.
- ICES. 2009b. Report of the Study Group on Biological Characteristics as Predictors of Salmon Abundance (SGBICEPS). ICES CM 2009/DFC: 02, 119 pp.
- ICES. 2010a. Report of the Working Group on North Atlantic Salmon (WGNAS), 22–31 March 2010 Copenhagen, Denmark. ICES CM 2010/ACOM: 09, 302 pp.
- ICES. 2010b. Report of the Study Group on Biological Characteristics as Predictors of Salmon Abundance (SGBICEPS), 24–26 November 2009, ICES Headquarters, Copenhagen, Denmark. ICES CM 2010/SSGEF: 03, 158 pp.
- ICES. 2011. Report of the Working Group on North Atlantic Salmon (WGNAS), 22–31 March 2011 Copenhagen, Denmark. ICES CM 2011/ACOM: 09, 286 pp.
- ICES. 2012a. Report of the Working Group on North Atlantic Salmon (WGNAS), 26 March–4 April 2012, Copenhagen, Denmark. ICES CM 2012/ACOM: 09, 322 pp.
- ICES. 2012b. ICES Advice 2012, Book 10, 99 pp.
- ICES. 2012c. Report of the Workshop on Eel and Salmon DCF Data (WKESDCF). 3–6 July 2012, Copenhagen, Denmark. ICES CM / ACOM: 62, 67 pp.
- ICES. 2013. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 3–12 April 2013. ICES CM 2013/ACOM: 09, 379 pp.
- ICES. 2014. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 19–28 March 2014. ICES CM 2014/ACOM: 09, 431 pp.
- IPCC. 2001. IPCC third assessment report: climate change 2001. Synthesis report (ed. R.T. Watson and the Core Writing Team). IPCC, Geneva, Switzerland, 184 pp.
- Jákupsstovu, S.H.I. 1988. Exploitation and migration of salmon in Faroese waters. In *Atlantic Salmon: Planning for the Future*, pp. 458–482. Ed. by D. Mills and D. Piggins. Croom Helm, London. 587 pp.
- Jacobsen, J.A., Lund, R.A., Hansen, L.P. and O'Maoléidigh, N. 2001. Seasonal differences in the origin of Atlantic salmon (*Salmo salar* L.) in the Norwegian Sea based on estimates from age structures and tag recaptures. *Fisheries Research* 52: 169–177.
- Jonsson, B. and Jonsson, N. 2009. A review of the likely effects of climate change on anadromous Atlantic salmon *Salmo salar* and brown trout *Salmo trutta*, with particular reference to water temperature and flow. *Journal of Fish Biology* 75: 2381–2447.
- Jutila, E., Jokikokko, E. and Julkunen, M. 2006. Long-term changes in the smolt size and age of Atlantic salmon, *Salmo salar* L., in a northern Baltic river related to parr density, growth opportunity and post-smolt survival. *Ecology of Freshwater Fish* 15: 321–330.
- Kennedy, R.J. and Crozier, W.W. 2010. Evidence of changing migratory patterns of wild Atlantic salmon *Salmo salar* smolts in the River Bush, Northern Ireland, and possible associations with climate change. *Journal of Fish Biology* 76, 1786–1805.
- Krkošek, M., Revie, C.W., Gargan, P.G., Skilbrei, O.T., Finstad, B., and Todd, C.D. 2013. Impact of parasites on salmon recruitment in the Northeast Atlantic Ocean. *Proceedings of the Royal Society B: Biological Sciences* 280: 20122359. <http://dx.doi.org/10.1098/rspb.2012.2359>

- Lunn, D.J., Thomas, A., Best, N. and Spiegelhalter, D. 2000. WinBUGS-A Bayesian modelling framework: Concepts, structure, and extensibility. *Statistics and Computing* 10: 325–337.
- MAFF. 1991. Salmon Net Fisheries: Report of a review of salmon net fishing in the areas of the Yorkshire and Northumbria regions of the National Rivers Authority and the salmon fishery districts from the River Tweed to the River Ugie. Pub. HMSO, London.
- Marine Scotland Science. 2010. The North Esk: Scotland's Monitored Salmon River. Topic Sheet 39. URL <http://www.scotland.gov.uk/Resource/Doc/295194/0099740.pdf>.
- Marine Scotland Science. 2012. Collecting the Marine Scotland salmon and sea trout fishery statistics. Topic Sheet 67. URL <http://www.scotland.gov.uk/Resource/0040/00402132.pdf>.
- Marine Scotland Science. 2013. Salmon Fishery Statistics. 2012 Season. Topic Sheet 68. URL <http://www.scotland.gov.uk/Resource/0043/00434129.pdf>.
- Michielsens, C.G.J., McAllister, M.K., Kuikka, S., Mäntyniemi, S., Romakkaniemi, A., Pakarinen, T., Karlsson, L. and Uusitalo, L. 2008. Combining multiple Bayesian data analyses in a sequential framework for quantitative fisheries stock assessment. *Canadian Journal of Fisheries and Aquatic Sciences* 65, 962–974.
- NASCO. 1994. Explanatory note on the 1994 West Greenland Quota Calculation. NASCO WGC(94), 4 pp.
- NASCO. 1998. North Atlantic Salmon Conservation Organisation. Agreement on the adoption of a precautionary approach. Report of the 15th annual meeting of the Council. CNL(98)46, 4 pp.
- NASCO. 1999. North Atlantic Salmon Conservation Organisation. Action plan for the application of the precautionary approach. CNL(99)48. 14 pp.
- NASCO. 2001. Report of the eighteenth annual meeting of the Commissions, Annex 14. NASCO, Edinburgh. pp. 183–191.
- Nicieza, A.G. and Braña, F. 1993. Relationships among smolt size, marine growth, and sea age at maturity of Atlantic salmon (*Salmo salar*) in northern Spain. *Canadian Journal of Fisheries and Aquatic Sciences* 50: 1632–1640.
- Niemelä, E., Erkinaro, J., Julkunen, M., Hassinen, E., Lämsman, M. and Brørs, S. 2006. Temporal variation in abundance, return rate and life histories of previously spawned Atlantic salmon in a large Sub-arctic river. *Journal of Fish Biology* 68: 1222–1240.
- O'Connell, M.F., Reddin, D.G., Amiro, P.G., Marshall, T.L., Chaput, G., Mullins, C.C., Locke, A., O'Neil, S.F. and Cairns, D.K. 1997. Estimates of conservation spawner requirements for Atlantic salmon (*Salmo salar* L.) for Canada. Department of Fisheries and Oceans Canadian Stock Assessment Secretariat Research Document 97/100, 58pp.
- O'Connell, M.F., Cochrane, N.M., Ash, E.G.M. and Mullins, C.C. 1998. An analysis of the License Stub Return System in the Newfoundland Region, 1994–1997. DFO, CSAS Res. Doc. 98/111, 67 p.
- Otero, J., L'Abée-Lund, J.H., Castro-Santos, T., Leonardsson, K., Størvik, G.O., Jonsson, B., Dempson, J.B., Russell, I.C., Jensen, A.J., Baglinière, J.-L., Dionne, M., Armstrong, J.D., Romakkaniemi, A., Letcher, B.H., Kocik, J.F., Erkinaro, J., Poole, R., Rogan, G., Lundqvist, H., Maclean, J.C., Jokikokko, E., Arnekleiv, J.V., Kennedy, R.J., Niemelä, E., Caballero, P., Music, P.A., Antonsson, T., Gudjonsson, S., Veselov, A.E., Lamberg, A., Groom, S., Taylor, B.H., Taberner, M., Dillane, M., Arnason, F., Horton, G., Hvidsten, N.A., Jonsson, I.R., Jonsson, N., McKelvey, S., Næsje, T.F., Skaala, O., Smith, G.W., Sægrov, H., Stenseth, N.C. and Vøllestad, L.A. 2014. Basin-scale phenology and effects of climate variability on global timing of initial seaward migration of Atlantic salmon (*Salmo salar*). *Global Change Biology* doi:10.1111/gcb.12363.
- Parrish, D.L., Behnke, R.J., Gephard, S.R., McCormick, S.D. and Reeves, G.H. 1998. Why aren't there more Atlantic salmon (*Salmo salar*)? *Canadian Journal of Fisheries and Aquatic Sciences* 55 (Suppl. 1): 281–287.

- Peyronnet, A., Friedland, K.D., O'Maoiléidigh, N., Manning, M. and Poole, W.R. 2007. Links between patterns of marine growth and survival of Atlantic salmon *Salmo salar*, L. *Journal of Fish Biology* 71: 684–700.
- Peyronnet, A., Friedland, K.D. and O'Maoiléidigh, N. 2008. Different ocean and climate factors control the marine survival of wild and hatchery Atlantic salmon *Salmo salar* in the north-east Atlantic Ocean. *Journal of Fish Biology* 73: 945–962.
- Potter, E.C.E. and Dunkley, D.A. 1993. Evaluation of marine exploitation of salmon in Europe. In *Salmon in the sea, and new enhancement strategies*. Edited by D. Mills. Fishing News Books, Blackwell Science, Oxford. pp. 203–219.
- Potter, E.C.E. 1996. Increases in the returns of salmon to homewaters following the reduction in fishing at Faroes and Greenland. In *Enhancement of Spring Salmon*. Edited by D. Mills. Proceedings of Atlantic Salmon Trust Conference, London, January 1996. Atlantic Salmon Trust, Pitlochry. pp. 67–86.
- Potter, E.C.E., Hansen, L.P., Gudbergsson, G., Crozier, W.W., Erkinaro, J., Insulander, C., MacLean, J., O'Maoileidigh, N. and Prusov, S. 1998. A method for estimating preliminary conservation limits for salmon stocks in the NASCO-NEAC area. International Council for the Exploration of the Sea CM 1998/T: 17. 11 pp.
- Potter, E.C.E. 2001. Past and present use of reference points for Atlantic salmon. In: *Stock, Recruitment and Reference Points: Assessment and Management of Atlantic Salmon*, pp. 195–223. Ed. by E. Prévost and G. Chaput. INRA Editions, Paris.
- Potter, E.C.E., and Hansen, L.P. 2001. Do Farm Escapees Distort Estimates Of Salmon Pre-Fishery Abundance And National Conservation Limits? ICES CM 2001/M:05.
- Potter, E.C.E., Ó Maoiléidigh, N. and Chaput, G. 2003. Marine mortality of Atlantic salmon, *Salmo salar* L: methods and measures. DFO Can. Sci. Adv. Sec. Res. Doc. 2003/101.
- Potter, E.C.E., Crozier, W.W., Schön, P.-J., Nicholson, M.D., Prévost, E., Erkinaro, J., Gudbergsson, G., Karlsson, L., Hansen, L.P., MacLean, J.C., Ó Maoiléidigh, N., and Prusov, S. 2004. Estimating and forecasting pre-fishery abundance of Atlantic salmon (*Salmo salar* L.) in the Northeast Atlantic for the management of mixed stock fisheries. *ICES Journal of Marine Science* 61: 1359–1369.
- Prairie, Y.T. 1996. Evaluating the predictive power of regression models. *Canadian Journal of Fisheries and Aquatic Sciences* 53: 490–492.
- R Development Core Team. 2007. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. ISBN 3-900051-07-0, URL <http://www.R-project.org>.
- Rago, P.J., Meerburg, D.J., Reddin, D.G., Chaput, G.J., Marshal, T.L., Dempson, B., Caron, F., Porter, T.R., Friedland, K.D. and Baum, E.T. 1993a. A continental run reconstruction model for the non-maturing component of North American Atlantic salmon: analysis of fisheries in Greenland and Newfoundland Labrador, 1974–1991. ICES CM 1993/M: 24, 21 pp.
- Rago, P.J., Reddin, D.G., Porter, T.R., Meerburg, D.J., Friedland, K.D. and Potter, E.C.E. 1993b. Estimation and analysis of pre-fishery abundance of the two-sea-winter populations of North American Atlantic salmon (*Salmo salar* L.), 1974–1991. ICES CM 1993/M: 25, 33 pp.
- Reddin, D.G. 2010. Atlantic salmon return and spawner estimates for Labrador. DFO Can. Sci. Advis. Section Res. Doc. 2009/045. iv + 19 p.
- Reddin, D.G. and Dempson, J.B. 1989. Harvest estimates of Sand Hill River Atlantic salmon. Canadian Atlantic Fisheries Scientific Advisory Committee Working Paper 89/104. 19 p.
- Reddin, D.G. and Veinott, G.I. 2010. Atlantic salmon return and spawner estimates for Newfoundland. DFO Can. Sci. Advis. Sec. Res. Doc. 2009/044. iv + 28 p.
- Reddin, D.G., Short, P.B., O'Connell, M.F. and Walsh, A.D. 1996. Atlantic salmon stock status for Sand Hill River, Labrador, 1995. DFO, CSAS Res. Doc. 96/82. 32 p.

- Reddin, D.G., Short, P.B. and Johnson, R.W. 1998. The stock status of Atlantic salmon (*Salmo salar* L.) of Big Brook (Michaels River), Labrador in 1997. DFO, CSAS Res. Doc. 98/119.
- Reddin, D.G., Anthony, R., Watts, K., Nuna, R. and Luther, R.J. 2004. Environmental conditions and harvests in various fisheries for salmonids in Labrador, 2002. CSAS Res. Doc. 2004/003, 20 p.
- Rosenthal, R. 1984. Meta-analytic procedures for social research. Sage Publications, London.
- Royce, W.F. 1984. Introduction to the practice of fishery science. Academic Press, New York.
- Russell, I.C., Aprahamian, M.W., Barry, J., Davidson, I.C., Fiske, P., Ibbotson, A.T., Kennedy, R.J., Maclean, J.C., Moore, A., Otero, J., Potter, E.C.E. and Todd, C.D. 2012. The influence of the freshwater environment and the biological characteristics of Atlantic salmon smolts on their subsequent marine survival. *ICES Journal of Marine Science* 69: 1563–1573.
- Shearer, W.M. 1992. The Atlantic salmon-Natural History, Exploitation and Future Management. Fishing News Books, Oxford, 244 pp.
- Skilbrei, O.T., Finstad, B., Urdal, K., Bakke, G., Kroglund, F. and Strand, R. 2013. Impact of early salmon louse, *Lepeophtheirus salmonis*, infestation and differences in survival and marine growth of sea-ranched Atlantic salmon, *Salmo salar* L., smolts 1997–2009. *Journal of Fish Diseases* 36: 249–260. doi:10.1111/jfd.12052.
- Todd, C.D., Hughes, S.L., Marshall, C.T., MacLean, J.C., Lonergan, M.E. and Biuw, E.M. 2008. Detrimental effects of recent ocean surface warming on growth condition of Atlantic salmon. *Global Change Biology* 14: 958–970.
- Vøllestad, L.A., Hirst, D., L'Abée-Lund, J.H., Armstrong, J.D., MacLean, J.C., Youngson, A.F. and Stenseth, N.C. 2009. Divergent trends in anadromous salmonid populations in Norwegian and Scottish rivers. *Proceedings of the Royal Society B-Biological Sciences* 276: 1021–1027.
- Vollset, K.W., Barlaup, B.T., Skoglund, H., Normann, E.S. and Skilbrei, O.T. 2014. Salmon lice increase the age of returning Atlantic salmon. *Biology Letters* 10. DOI: 10.1098/rsbl.2013.0896.
- Youngson, A.F. and Webb, J.H. 1993. Thyroid hormone levels in Atlantic salmon (*Salmo salar*) during the return migration from the ocean to spawn. *Journal of Fish Biology* 42: 293–300.

Appendices to Stock Annex

Appendix 1 (a): Description of how catch-and-release mortality is incorporated in regional and national stock assessments

| Commission Area | Country/Region | How it is used in regional and national assessments | Future developments / improvements |
|-----------------|--------------------------------|--------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|-------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| NAC | Canada-Quebec | C&R has become more popular in the region and C&R only angling licenses are sold. C&R data are incomplete as there is no requirement to report C&R numbers. Generally, C&R mortality is considered in the assessment but the majority of the assessments are conducted as spawner counts after the fisheries so any losses due to C&R mortality are accounted for in the spawner estimates but not in the returns (which are the sum of known losses and spawning escapement). | New studies of the contribution of C&R fish to spawning success have been initiated. C&R monitoring is becoming more complete. Consideration will be given in the future to incorporating these losses in the returns and in the assessments based on angling catches, especially as reporting improves. |
| | Canada-Newfoundland & Labrador | Catch and release mortality is included in estimates of spawners. Spawning escapement is reduced by 5-15% (mean 10%) of the released catch. | No plans for further development. |
| | Canada - Gulf | Assessments of spawners are adjusted by mortality rates of 3% to 6% of the total C&R estimates of small and large salmon. The rates vary by river according to angling seasons, and the occurrence of other factors such as disease which can affect survival of salmon. | Catch and release mortality is known to be affected by the water temperatures when fish are angled. In some cases, angling fisheries are closed when water temperatures are high in the summer to reduce the losses of fish from C&R. Methods to determine catch and release numbers vary by river and in some cases, the number of released fish is estimated from returns and historical creel survey data. As the practice of C&R becomes more popular, estimation methods for C&R values will have to be revisited. |
| | Canada – Scotia/Fundy | Assessments are currently adjusted by 4% of the C&R fish to correct for C&R mortality. | Numbers of C&R fish are currently low (retention fisheries are closed). If C&R catches increase, further research on the correction factor would be warranted. |
| | USA | No correction for mortality due to C&R used in estimating spawner numbers. However, all fisheries have been closed and the number of fish caught relative to stock size is very small. | |
| NEAC | Russia | With increasing C&R the retained catch for similar effort is reduced. Therefore the exploitation rate for retained fish is lower. The increase in C&R in recent years is incorporated into the national run-reconstruction model by reducing the exploitation rate value used in the model input. This is assessed qualitatively. No correction for increased C&R mortality is applied when estimating the spawning escapement. | If C&R information is incorporated into formal assessments then multiple recaptures should be taken into account. C&R mortality should be incorporated into estimates of spawning escapement. |
| | Norway | | |
| | Sweden | | |
| | Iceland | | |
| | UK(Scotland) | | |
| | Ireland | No correction for mortality due to C&R used in estimating spawner numbers or in the national run-reconstruction model. | Incorporation of formal method for estimating the effect of C&R on number of returning fish. Incorporation of C&R mortality in estimates of spawning escapement |
| | UK(England & Wales) | With increasing C&R the retained catch for similar effort is reduced. Therefore the exploitation rate for retained fish is lower. The increase in C&R in recent years is incorporated into the national run-reconstruction model by reducing the exploitation rate value used in the model input. This is assessed qualitatively. 20% mortality of C&R fish used in assessing compliance with river-specific conservation | If C&R information is incorporated into formal assessments then multiple recaptures should be taken into account. |
| | UK(N. Ireland) | Returns are estimated by raising the reported net catch by exploitation rate. No correction for increased C&R mortality is applied when estimating the spawning escapement. | If C&R information is incorporated into formal assessments then multiple recaptures should be taken into account. C&R mortality should be incorporated into estimates of spawning escapement. |
| | Denmark | C&R rates recorded, but no national run-reconstruction assessment applied. | |
| | Finland | No record of C&R | If C&R information is collected, it should be incorporated into formal assessments and multiple recaptures should be |

Appendix 1 (b): Description of how unreported catch is incorporated in regional, national and international stock assessments

| Commission Area | Country/Region | How it is used in regional and national assessments | How used in international assessments | Future developments / improvements |
|-----------------|--------------------------------|------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| NAC | Canada-Quebec | Unreported catches are based on historical estimates relative to stock size or are provided by field conservation and protection staff. Unreported catches when available are included in the regional assessment of returns and spawners. | Unreported catches which occur in marine waters outside the jurisdiction of the regions are not included in the run reconstruction models. | If unreported catch estimates were provided they could be incorporated in the regional assessments and in the continent estimates of abundance and spawners. Unreported catch may be accounted for in either the returns or the spawners, depending upon when and where the illegal activity occurs relative to the location and time of the assessment model. |
| | Canada-Newfoundland & Labrador | Catch statistics include estimates of harvests by log book non-respondents. Therefore they are included in the regional assessments and the PFA estimate. No account is taken of illegal fisheries. | | |
| | Canada - Gulf | Unreported catches are sometimes provided by Conservation and Protection Personnel and are estimates of illegal fishing removals within specific regions. Unreported catches have not been used in the assessments of returns or spawners. | | |
| | Canada – Scotia/Fundy | No adjustment made, with the exception of the Saint John River where returns/spawners are adjusted for estimated bycatch and poaching. In other rivers where assessments directly quantify spawners, returns would be underestimated if catch is under reported. | | |
| | USA | Unreported catch is estimated to be zero and therefore has no effect on national assessments. | | |
| NEAC | Russia | Minimum and maximum estimates of the unreporting rate are used in deriving national PFA estimates from the catch of 1SW & MSW salmon. | National estimates (which incorporate unreported catches) are aggregated to provide PFA, return and spawner estimates for stock complexes. | Incorporate revised estimates of minimum and maximum estimates of unreporting rate as national estimates are improved. |
| | Finland | | | |
| | Norway | | | |
| | Sweden | | | |
| | Iceland | | | |
| | Ireland | | | |
| | UK(Scotland) | | | |
| | UK(England & Wales) | | | |
| | UK(N. Ireland) | | | |
| | France | | | |
| NEAC | Denmark | No national assessment | Assumed to be negligible unreported catch. Estimate of discard mortality for 1SW fish is incorporated in stock assessments. | Sampling programme if fishery resumes. |
| | Faroes | Not applicable | | |
| W. Greenland | W. Greenland | Not applicable | Unreported catch at West Greenland is incorporated in assessments for both the NAC and NEAC areas. Since 1993, this has been provided by the Greenlandic authorities. Prior to this time, no unreported catch component is included in the models. | Annual variation in unreported catch estimates would be incorporated into the model. |

Appendix 2: Overview of current DCF and future data needs for Atlantic salmon assessment/ scientific advice

| TYPE OF DATA | COLLECTED UNDER DCF | AVAILABLE TO WG | REVIEWED AND EVALUATED BY WG | USED IN CURRENT ASSESSMENT MODELS | FUTURE PLANS | NOTES |
|-------------------------|--------------------------|--------------------------|------------------------------|-----------------------------------------------|------------------------------------------------------------------------------------------------|---------------------------------------------------------------------------------------------------------------------------------------------------------|
| How to be filled | Yes/ No/ Partially | Yes/ No/ Partially | Yes/ No/ Partially | Yes/ No/ Partially used | Keep as current DCF/ Improve sampling intensity/ No need to be collected/ (other free text) | Free text |
| Fleet capacity | No ** | No * | No | No | No need to be collected | See 'Fishing gear and effort' |
| Fuel consumption | No ** | No * | No | No | No need to be collected | Many salmon fisheries use unpowered vessels |
| Fishing gear and effort | Partially ** | Partially | Partially | Partially, but information requested by NASCO | Use for estimation of exploitation rates. Improve coverage and sampling intensity in DC-MAP | Data required for all relevant areas/fisheries |
| Landings | Partially ** | Yes | Yes | Yes | Improve coverage in DC-MAP | Data required on: catch in numbers and weights for recreational and commercial fisheries in rivers, estuaries and coastal waters. |
| Discards | No ** | No * | No | No | No need to be collected | Not relevant to salmon except (historically) in Faroes fishery. NB: 'catch and release' fish are deliberately caught and so not classed as discards. |

| TYPE OF DATA | COLLECTED UNDER DCF | AVAILABLE TO WG | REVIEWED AND EVALUATED BY WG | USED IN CURRENT ASSESSMENT MODELS | FUTURE PLANS | NOTES |
|------------------------------------|--------------------------------------------------------------|------------------------------|------------------------------|-------------------------------------------------------------------|---------------------------------------------------|------------------------------------------------------------------------------------------------------------|
| Recreational fisheries | Partially ** | Yes | Yes | Yes | Improve coverage in DC-MAP | Extent of DCF coverage unclear. Complete catch data needed for all recreational fisheries (see 'Landings') |
| Catch & Release | No ** | Partially | Partially | No - but data requested by NASCO | Include collection in DC-MAP | Data on numbers of fish caught and released required for all recreational fisheries |
| cpue dataserries | Partially ** | Partially | Partially | Partially | Improve sampling intensity in DC-MAP | Data used to generate national inputs to models |
| Age composition | Partially ** Some ageing based on fish lengths or weights | Yes | Yes | Yes | Improve coverage and sampling intensity in DC-MAP | Extent of DCF coverage unclear; sampling intensities in other fisheries inappropriate to salmon |
| Wild/reared origin (scale reading) | No ** | Partially from other sources | Partially | Partially used - information on farmed fish is requested by NASCO | Improve sampling intensity in DC-MAP | Extent of DCF coverage unclear |
| Length and weight-at-age | Partially ** | Partially | Yes | Yes - but some ageing based on fish lengths or weights | Improve sampling coverage in DC-MAP | DCF does not cover all relevant areas/fisheries; sampling intensities inappropriate to salmon |
| Sex ratio | No ** | Yes- from other sources | Partially | Yes | Modify sampling intensity in DC-MAP | Estimates required at national/regional level every five years |

| TYPE OF DATA | COLLECTED UNDER DCF | AVAILABLE TO WG | REVIEWED AND EVALUATED BY WG | USED IN CURRENT ASSESSMENT MODELS | FUTURE PLANS | NOTES |
|---------------------------------------------------------------------|-----------------------------------------------------------|-----------------|------------------------------|-----------------------------------|-----------------------------------------------------------|------------------------------------------------------------------------------------------------------------|
| Maturity | Not known ** | No * | No | No | No need to be collected – all returning adults are mature | DCF requires collection but extent of coverage unclear; data not required for assessments |
| Fecundity | No ** | Yes | Partially | Yes | Include collection in DC-MAP | Estimates required at national/regional level every 5 years |
| Data processing industry | No ** | No ** | No | No | No need to be collected | Requirement not clear |
| Juvenile surveys (Electrofishing) | Partially ** but not requested for Atlantic salmon in DCF | Yes | Partially | Partially | Include collection in DC-MAP | Data used to develop reference points and confirm stock status. Also required for assessments under WFD |
| Adult census data (Counters, fish ladders, etc.) | Partially ** but not requested for Atlantic salmon in DCF | Yes | Partially | Yes | Include collection in DC-MAP | Counts required for ~one river in 30. Data required to provide exploitation rates for assessments |
| Index river data (Smolt & adult trapping; tagging programmes; etc.) | Partially ** but not requested for Atlantic salmon in DCF | Yes | Partially | Yes | Include collection in DC-MAP | Index rivers are identified by ICES. Data used to develop reference points and inputs to assessment models |
| Genetic data (for mixed- | No ** | Partially | Partially - | Not currently | Include collection in DC- | Genetic analysis |

| TYPE OF DATA | COLLECTED UNDER DCF | AVAILABLE TO WG | REVIEWED AND EVALUATED BY WG | USED IN CURRENT ASSESSMENT MODELS | FUTURE PLANS | NOTES |
|------------------|---------------------|----------------------------------------------|--------------------------------|---------------------------------------------------------------|----------------------------------------------------------|------------------------------------------------------------------------------------|
| stock analysis) | | | for some mixed-stock fisheries | | MAP - sampling in mixed-stock fisheries every five years | is now advised to provide more reliable stock composition in mixed-stock fisheries |
| Economic data | Not known ** | No * | No | No - but data are of use to NASCO | | Collection of economic data would be useful to managers |
| Aquaculture data | Not known ** | Partially - marine farm production collected | Yes | No - but information on farm production is requested by NASCO | | Currently not required for freshwater |

Add other data type to the cells with a light blue shading, if needed.

* Not asked for by the ICES WGNAS.

**) Not mandatory for some or all areas/stocks/fisheries under the current DCF.

Appendix 3: Input data for NEAC Pre Fishery Abundance analysis using Monte Carlo simulation

Finland

| Annual input data for NEAC PFA run-reconstruction & NCL models for FINLAND (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation) | | | | | | | | | | |
|-----------------------------------------------------------------------------------------------------------------------------------------------------------------------------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
| 1971 | 8,422 | 8,538 | 35.0 | 5.0 | 35.0 | 5.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1972 | 13,160 | 13,341 | 35.0 | 5.0 | 35.0 | 5.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1973 | 11,969 | 15,958 | 35.0 | 5.0 | 35.0 | 5.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1974 | 23,709 | 23,709 | 35.0 | 5.0 | 35.0 | 5.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1975 | 16,527 | 26,417 | 35.0 | 5.0 | 35.0 | 5.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1976 | 11,323 | 21,719 | 35.0 | 5.0 | 35.0 | 5.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1977 | 5,807 | 13,227 | 35.0 | 5.0 | 35.0 | 5.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1978 | 7,902 | 8,452 | 35.0 | 5.0 | 35.0 | 5.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1979 | 9,249 | 7,390 | 35.0 | 5.0 | 35.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1980 | 4,792 | 8,938 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1981 | 7,386 | 9,835 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1982 | 2,163 | 12,826 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1983 | 10,680 | 13,990 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1984 | 11,942 | 13,262 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1985 | 18,039 | 10,339 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1986 | 16,389 | 9,028 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1987 | 20,950 | 11,290 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1988 | 10,019 | 7,231 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1989 | 28,091 | 10,011 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1990 | 26,646 | 12,562 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1991 | 32,423 | 15,136 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1992 | 42,965 | 16,158 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1993 | 30,197 | 18,720 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1994 | 12,016 | 15,521 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1995 | 11,801 | 9,634 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1996 | 22,799 | 6,956 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1997 | 19,481 | 10,083 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1998 | 22,460 | 8,497 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1999 | 38,687 | 8,854 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 10.0 | 50.0 | 10.0 |
| 2000 | 40,730 | 21,088 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 10.0 | 50.0 | 10.0 |
| 2001 | 29,501 | 28,112 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 10.0 | 50.0 | 10.0 |
| 2002 | 16,721 | 24,642 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2003 | 16,497 | 17,751 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2004 | 7,002 | 8,062 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2005 | 15,366 | 6,685 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2006 | 26,916 | 10,533 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2007 | 7,862 | 15,269 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2008 | 8,481 | 15,355 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2009 | 15,042 | 6,587 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2010 | 12,085 | 10,590 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2011 | 13,727 | 8,152 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2012 | 23,764 | 9,851 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2013 | 13,724 | 9,494 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2014 | 19,495 | 10,302 | 25.0 | 5.0 | 25.0 | 5.0 | 50.0 | 10.0 | 50.0 | 10.0 |

France

Annual input data for NEAC PFA run-reconstruction & NCL models for FRANCE. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | 1,740 | 4,060 | NA | NA | NA | NA | 3.5 | 1.5 | 37.5 | 12.5 |
| 1972 | 3,480 | 8,120 | NA | NA | NA | NA | 3.5 | 1.5 | 37.5 | 12.5 |
| 1973 | 2,130 | 4,970 | NA | NA | NA | NA | 3.5 | 1.5 | 37.5 | 12.5 |
| 1974 | 990 | 2,310 | NA | NA | NA | NA | 3.5 | 1.5 | 37.5 | 12.5 |
| 1975 | 1,980 | 4,620 | NA | NA | NA | NA | 3.5 | 1.5 | 37.5 | 12.5 |
| 1976 | 1,820 | 3,380 | NA | NA | NA | NA | 3.5 | 1.5 | 37.5 | 12.5 |
| 1977 | 1,400 | 2,600 | NA | NA | NA | NA | 3.5 | 1.5 | 37.5 | 12.5 |
| 1978 | 1,435 | 2,665 | NA | NA | NA | NA | 3.5 | 1.5 | 37.5 | 12.5 |
| 1979 | 1,645 | 3,055 | NA | NA | NA | NA | 3.5 | 1.5 | 37.5 | 12.5 |
| 1980 | 3,430 | 6,370 | NA | NA | NA | NA | 3.5 | 1.5 | 37.5 | 12.5 |
| 1981 | 2,720 | 4,080 | NA | NA | NA | NA | 3.5 | 1.5 | 35.0 | 15.0 |
| 1982 | 1,680 | 2,520 | NA | NA | NA | NA | 3.5 | 1.5 | 35.0 | 15.0 |
| 1983 | 1,800 | 2,700 | NA | NA | NA | NA | 3.5 | 1.5 | 35.0 | 15.0 |
| 1984 | 2,960 | 4,440 | NA | NA | NA | NA | 3.5 | 1.5 | 35.0 | 15.0 |
| 1985 | 1,100 | 3,330 | NA | NA | NA | NA | 3.5 | 1.5 | 35.0 | 15.0 |
| 1986 | 3,400 | 3,400 | NA | NA | NA | NA | 7.0 | 5.0 | 35.0 | 15.0 |
| 1987 | 6,013 | 1,806 | NA | NA | NA | NA | 7.0 | 5.0 | 35.0 | 15.0 |
| 1988 | 2,063 | 4,964 | NA | NA | NA | NA | 7.0 | 5.0 | 35.0 | 15.0 |
| 1989 | 1,124 | 2,282 | NA | NA | NA | NA | 7.0 | 5.0 | 35.0 | 15.0 |
| 1990 | 1,886 | 2,332 | NA | NA | NA | NA | 7.0 | 5.0 | 35.0 | 15.0 |
| 1991 | 1,362 | 2,125 | NA | NA | NA | NA | 7.0 | 5.0 | 35.0 | 15.0 |
| 1992 | 2,490 | 2,671 | NA | NA | NA | NA | 7.0 | 5.0 | 35.0 | 15.0 |
| 1993 | 3,581 | 1,254 | NA | NA | NA | NA | 7.0 | 5.0 | 35.0 | 15.0 |
| 1994 | 2,810 | 2,290 | NA | NA | NA | NA | 7.0 | 5.0 | 30.0 | 10.0 |
| 1995 | 1,669 | 1,095 | NA | NA | NA | NA | 12.5 | 7.5 | 30.0 | 10.0 |
| 1996 | 2,063 | 1,943 | NA | NA | NA | NA | 12.5 | 7.5 | 30.0 | 10.0 |
| 1997 | 1,060 | 1,001 | NA | NA | NA | NA | 12.5 | 7.5 | 30.0 | 10.0 |
| 1998 | 2,065 | 846 | NA | NA | NA | NA | 12.5 | 7.5 | 30.0 | 10.0 |
| 1999 | 690 | 1,831 | NA | NA | NA | NA | 12.5 | 7.5 | 30.0 | 10.0 |
| 2000 | 1,792 | 1,277 | NA | NA | NA | NA | 12.5 | 7.5 | 30.0 | 10.0 |
| 2001 | 1,544 | 1,489 | NA | NA | NA | NA | 12.5 | 7.5 | 30.0 | 10.0 |
| 2002 | 2,423 | 1,065 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2003 | 1,598 | 1,540 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2004 | 1,927 | 2,880 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2005 | 1,256 | 1,771 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2006 | 1,763 | 1,785 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2007 | 1,378 | 1,685 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2008 | 1,365 | 1,865 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2009 | 389 | 863 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2010 | 1,313 | 711 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2011 | 899 | 1,998 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2012 | 974 | 1,585 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2013 | 1,371 | 1,632 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |
| 2014 | 1,217 | 2,027 | 30.0 | 10.0 | 22.5 | 7.5 | 12.5 | 7.5 | 30.0 | 10.0 |

Iceland (South and West)

Annual input data for NEAC PFA run-reconstruction & NCL models for ICELAND (SOUTH-WEST). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | 30,618 | 16,749 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1972 | 24,832 | 25,733 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1973 | 26,624 | 23,183 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1974 | 18,975 | 20,017 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1975 | 29,428 | 21,266 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1976 | 23,233 | 18,379 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1977 | 23,802 | 17,919 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1978 | 31,199 | 23,182 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1979 | 28,790 | 14,840 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1980 | 13,073 | 20,855 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1981 | 16,890 | 13,919 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1982 | 17,331 | 9,826 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1983 | 21,923 | 16,423 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1984 | 13,476 | 13,923 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1985 | 21,822 | 10,097 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1986 | 35,891 | 8,423 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1987 | 22,302 | 7,480 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1988 | 40,028 | 8,523 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1989 | 22,377 | 7,607 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1990 | 20,584 | 7,548 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1991 | 22,711 | 7,519 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1992 | 26,006 | 8,479 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1993 | 25,479 | 4,155 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1994 | 20,985 | 6,736 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1995 | 25,371 | 6,777 | 4.0 | 1.0 | 4.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1996 | 21,913 | 4,364 | 4.0 | 1.0 | 4.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1997 | 16,007 | 4,910 | 4.0 | 1.0 | 4.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1998 | 21,900 | 3,037 | 4.0 | 1.0 | 4.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1999 | 17,448 | 5,757 | 4.0 | 1.0 | 4.0 | 1.0 | 49.0 | 10.0 | 68.0 | 10.0 |
| 2000 | 15,502 | 1,519 | 4.0 | 1.0 | 4.0 | 1.0 | 49.0 | 10.0 | 66.0 | 10.0 |
| 2001 | 13,586 | 2,707 | 4.0 | 1.0 | 4.0 | 1.0 | 48.0 | 10.0 | 67.0 | 10.0 |
| 2002 | 16,952 | 2,845 | 4.0 | 1.0 | 4.0 | 1.0 | 48.0 | 10.0 | 65.0 | 10.0 |
| 2003 | 20,271 | 4,751 | 4.0 | 1.0 | 4.0 | 1.0 | 48.0 | 10.0 | 68.0 | 10.0 |
| 2004 | 20,319 | 3,784 | 4.0 | 1.0 | 4.0 | 1.0 | 48.0 | 10.0 | 67.0 | 10.0 |
| 2005 | 29,969 | 3,241 | 4.0 | 1.0 | 4.0 | 1.0 | 48.0 | 10.0 | 65.0 | 10.0 |
| 2006 | 21,153 | 2,689 | 4.0 | 1.0 | 4.0 | 1.0 | 48.0 | 10.0 | 65.0 | 10.0 |
| 2007 | 23,728 | 1,679 | 4.0 | 1.0 | 4.0 | 1.0 | 47.0 | 9.0 | 66.0 | 10.0 |
| 2008 | 28,774 | 1,659 | 4.0 | 1.0 | 4.0 | 1.0 | 47.0 | 10.0 | 57.0 | 10.0 |
| 2009 | 33,190 | 2,838 | 4.0 | 1.0 | 4.0 | 1.0 | 48.0 | 10.0 | 63.0 | 10.0 |
| 2010 | 33,318 | 6,061 | 4.0 | 1.0 | 4.0 | 1.0 | 47.0 | 10.0 | 65.0 | 10.0 |
| 2011 | 23,436 | 2,934 | 4.0 | 1.0 | 4.0 | 1.0 | 47.0 | 10.0 | 62.0 | 10.0 |
| 2012 | 13,312 | 1,429 | 4.0 | 1.0 | 4.0 | 1.0 | 47.0 | 10.0 | 53.0 | 10.0 |
| 2013 | 39,637 | 4,105 | 4.0 | 1.0 | 4.0 | 1.0 | 47.0 | 10.0 | 55.0 | 10.0 |
| 2014 | 10,695 | 4,349 | 4.0 | 1.0 | 4.0 | 1.0 | 47.0 | 10.0 | 55.0 | 10.0 |

Iceland (North and East)

Annual input data for NEAC PFA run-reconstruction & NCL models for ICELAND (NORTH-EAST). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | 4,610 | 6,625 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1972 | 4,223 | 10,337 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1973 | 5,060 | 9,672 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1974 | 5,047 | 9,176 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1975 | 6,152 | 10,136 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1976 | 6,184 | 8,350 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1977 | 8,597 | 11,631 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1978 | 8,739 | 14,998 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1979 | 8,363 | 9,897 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1980 | 1,268 | 13,784 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1981 | 6,528 | 4,827 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1982 | 3,007 | 5,539 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1983 | 4,437 | 4,224 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1984 | 1,611 | 5,447 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1985 | 11,116 | 3,511 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1986 | 13,827 | 9,569 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1987 | 8,145 | 9,908 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1988 | 11,775 | 6,381 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1989 | 6,342 | 5,414 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1990 | 4,752 | 5,709 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1991 | 6,900 | 3,965 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1992 | 12,996 | 5,903 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1993 | 10,689 | 6,672 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1994 | 3,414 | 5,656 | 2.0 | 1.0 | 2.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1995 | 8,776 | 3,511 | 4.0 | 1.0 | 4.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1996 | 4,681 | 4,605 | 4.0 | 1.0 | 4.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1997 | 6,406 | 2,594 | 4.0 | 1.0 | 4.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1998 | 10,905 | 3,780 | 4.0 | 1.0 | 4.0 | 1.0 | 50.0 | 10.0 | 70.0 | 10.0 |
| 1999 | 5,326 | 4,030 | 4.0 | 1.0 | 4.0 | 1.0 | 48.0 | 10.0 | 65.0 | 10.0 |
| 2000 | 5,595 | 2,324 | 4.0 | 1.0 | 4.0 | 1.0 | 48.0 | 10.0 | 64.0 | 10.0 |
| 2001 | 4,976 | 2,587 | 4.0 | 1.0 | 4.0 | 1.0 | 47.0 | 10.0 | 62.0 | 10.0 |
| 2002 | 8,437 | 2,366 | 4.0 | 1.0 | 4.0 | 1.0 | 46.0 | 10.0 | 60.0 | 10.0 |
| 2003 | 4,478 | 2,194 | 4.0 | 1.0 | 4.0 | 1.0 | 46.0 | 10.0 | 53.0 | 10.0 |
| 2004 | 11,823 | 2,239 | 4.0 | 1.0 | 4.0 | 1.0 | 45.0 | 10.0 | 55.0 | 10.0 |
| 2005 | 10,297 | 2,726 | 4.0 | 1.0 | 4.0 | 1.0 | 44.0 | 10.0 | 54.0 | 10.0 |
| 2006 | 11,082 | 2,179 | 4.0 | 1.0 | 4.0 | 1.0 | 45.0 | 10.0 | 45.0 | 10.0 |
| 2007 | 8,046 | 1,672 | 4.0 | 1.0 | 4.0 | 1.0 | 44.0 | 10.0 | 36.0 | 10.0 |
| 2008 | 7,021 | 2,693 | 4.0 | 1.0 | 4.0 | 1.0 | 42.0 | 10.0 | 45.0 | 10.0 |
| 2009 | 10,779 | 1,735 | 4.0 | 1.0 | 4.0 | 1.0 | 40.0 | 10.0 | 36.0 | 10.0 |
| 2010 | 8,621 | 2,602 | 4.0 | 1.0 | 4.0 | 1.0 | 40.0 | 10.0 | 38.0 | 10.0 |
| 2011 | 6,759 | 2,596 | 4.0 | 1.0 | 4.0 | 1.0 | 38.0 | 10.0 | 34.0 | 10.0 |
| 2012 | 3,699 | 1,419 | 4.0 | 1.0 | 4.0 | 1.0 | 40.0 | 10.0 | 33.0 | 10.0 |
| 2013 | 8,375 | 1,528 | 4.0 | 1.0 | 4.0 | 1.0 | 38.0 | 10.0 | 31.0 | 10.0 |
| 2014 | 3,573 | 1,617 | 4.0 | 1.0 | 4.0 | 1.0 | 38.0 | 10.0 | 27.0 | 10.0 |

Ireland

| Annual input data for NEAC PFA run-reconstruction & NCL models for IRELAND. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation) | | | | | | | | | | | | | | | | | | |
|------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|-------------------------------|-------------------------------|------------------------------|------------------------------|----------------------------|----------------------------|-----------------------------|-----------------------------|
| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon | Declared net catch 1SW salmon | Declared net catch MSW salmon | Catch and release 1SW salmon | Catch and release MSW salmon | 1SW salmon in Small rivers | MSW salmon in Small rivers | 1SW salmon in closed rivers | MSW salmon in closed rivers |
| 1971 | 409,965 | 46,594 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1972 | 437,089 | 49,863 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1973 | 476,131 | 54,008 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1974 | 542,124 | 60,976 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1975 | 598,524 | 68,260 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1976 | 407,018 | 47,358 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1977 | 351,745 | 41,256 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1978 | 307,569 | 35,708 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1979 | 282,700 | 32,144 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1980 | 215,116 | 35,447 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1981 | 137,366 | 26,101 | 37.5 | 7.5 | 37.5 | 7.5 | 75.7 | 11.4 | 47.5 | 12.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1982 | 269,847 | 11,754 | 37.5 | 7.5 | 37.5 | 7.5 | 71.9 | 10.8 | 36.7 | 8.3 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1983 | 437,751 | 26,479 | 37.5 | 7.5 | 37.5 | 7.5 | 66.1 | 9.9 | 40.1 | 7.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1984 | 224,872 | 20,685 | 37.5 | 7.5 | 37.5 | 7.5 | 64.6 | 9.7 | 43.5 | 6.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1985 | 430,315 | 18,830 | 37.5 | 7.5 | 37.5 | 7.5 | 74.6 | 11.2 | 36.1 | 3.4 | NA | NA | NA | NA | NA | 961 | NA | NA |
| 1986 | 443,701 | 27,111 | 37.5 | 7.5 | 37.5 | 7.5 | 68.7 | 10.3 | 46.0 | 9.0 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1987 | 324,709 | 26,301 | 30.0 | 10.0 | 30.0 | 10.0 | 69.8 | 10.5 | 32.2 | 4.7 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1988 | 391,475 | 22,067 | 30.0 | 10.0 | 30.0 | 10.0 | 62.0 | 9.3 | 37.4 | 5.6 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1989 | 297,797 | 25,447 | 30.0 | 10.0 | 30.0 | 10.0 | 65.7 | 9.9 | 47.2 | 8.8 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1990 | 172,098 | 15,549 | 30.0 | 10.0 | 30.0 | 10.0 | 60.7 | 9.1 | 59.9 | 6.1 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1991 | 120,408 | 10,334 | 30.0 | 10.0 | 30.0 | 10.0 | 59.5 | 8.9 | 26.5 | 3.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1992 | 182,255 | 15,456 | 30.0 | 10.0 | 30.0 | 10.0 | 62.1 | 9.3 | 51.5 | 3.8 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1993 | 150,274 | 13,156 | 25.0 | 10.0 | 25.0 | 10.0 | 58.6 | 8.8 | 42.0 | 18.0 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1994 | 234,126 | 20,506 | 25.0 | 10.0 | 25.0 | 10.0 | 71.4 | 10.7 | 40.5 | 2.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1995 | 232,480 | 20,454 | 25.0 | 10.0 | 25.0 | 10.0 | 63.5 | 9.5 | 41.8 | 1.2 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1996 | 203,920 | 18,021 | 25.0 | 10.0 | 25.0 | 10.0 | 59.9 | 9.0 | 55.1 | 3.2 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1997 | 170,774 | 14,724 | 25.0 | 10.0 | 15.0 | 5.0 | 50.1 | 7.5 | 30.8 | 12.2 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1998 | 191,868 | 17,269 | 25.0 | 10.0 | 15.0 | 5.0 | 53.7 | 8.1 | 61.9 | 1.4 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1999 | 158,818 | 14,801 | 25.0 | 10.0 | 15.0 | 5.0 | 47.8 | 7.2 | 34.1 | 18.1 | NA | NA | NA | NA | NA | NA | NA | NA |
| 2000 | 199,827 | 16,848 | 25.0 | 10.0 | 15.0 | 5.0 | 43.2 | 6.5 | 31.0 | 4.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 2001 | 218,715 | 18,436 | 7.5 | 2.5 | 7.5 | 2.5 | 48.0 | 7.2 | 35.0 | 8.0 | NA | NA | NA | NA | NA | NA | NA | NA |
| 2002 | 198,719 | 16,702 | 7.5 | 2.5 | 7.5 | 2.5 | 49.9 | 7.5 | 27.5 | 7.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 2003 | 161,270 | 13,745 | 7.5 | 2.5 | 7.5 | 2.5 | 41.3 | 6.2 | 21.5 | 5.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 2004 | 142,251 | 12,299 | 7.5 | 2.5 | 7.5 | 2.5 | 49.5 | 7.5 | 35.0 | 8.0 | NA | NA | NA | NA | NA | NA | NA | NA |
| 2005 | 127,371 | 10,716 | 7.5 | 2.5 | 7.5 | 2.5 | 44.5 | 6.5 | 23.5 | 3.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 2006 | 101,938 | 9,740 | 7.5 | 2.5 | 7.5 | 2.5 | 46.5 | 6.5 | 29.5 | 13.5 | NA | NA | NA | NA | NA | NA | NA | NA |
| 2007 | 30,418 | 2,477 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 8,334 | 679 | 12,137 | 988 | 9,548 | 777 | 40,255 | 3,278 |
| 2008 | 30,257 | 3,935 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 8,253 | 650 | 10,485 | 1,492 | 12,206 | 961 | 34,382 | 4,580 |
| 2009 | 24,184 | 4,756 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 6,264 | 493 | 9,799 | 1,623 | - | - | 46,570 | 4,964 |
| 2010 | 33,211 | 3,297 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 13,125 | 1,034 | 13,903 | 1,255 | - | - | 35,804 | 1,504 |
| 2011 | 29,117 | 3,970 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 11,071 | 902 | 11,222 | 1,530 | - | - | 33,251 | 1,208 |
| 2012 | 29,979 | 4,198 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 9,542 | 777 | 10,429 | 1,463 | - | - | 32,964 | 1,648 |
| 2013 | 24,029 | 4,831 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 13,378 | 747 | 8,821 | 1,861 | - | - | 45,764 | 1,698 |
| 2014 | 13317 | 1,935 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 8,332 | 624 | 4,715 | 685 | - | - | 50,211 | 1,441 |

Norway (Southeast)

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (SOUTH-EAST). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1972 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1973 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1974 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1975 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1976 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1977 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1978 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1979 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1980 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1981 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1982 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1983 | 9,039 | 9,004 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1984 | 11,402 | 11,527 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1985 | 18,699 | 11,883 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1986 | 23,089 | 12,077 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1987 | 19,601 | 14,179 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1988 | 17,520 | 9,443 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1989 | 23,965 | 12,254 | 50.0 | 10.0 | 50.0 | 10.0 | 65.0 | 10.0 | 60.0 | 10.0 |
| 1990 | 25,792 | 11,502 | 50.0 | 10.0 | 50.0 | 10.0 | 65.0 | 10.0 | 60.0 | 10.0 |
| 1991 | 21,064 | 10,753 | 50.0 | 10.0 | 50.0 | 10.0 | 65.0 | 10.0 | 60.0 | 10.0 |
| 1992 | 26,044 | 15,332 | 50.0 | 10.0 | 50.0 | 10.0 | 65.0 | 10.0 | 60.0 | 10.0 |
| 1993 | 23,070 | 12,596 | 40.0 | 10.0 | 40.0 | 10.0 | 85.0 | 10.0 | 60.0 | 10.0 |
| 1994 | 23,987 | 9,988 | 40.0 | 10.0 | 40.0 | 10.0 | 65.0 | 10.0 | 60.0 | 10.0 |
| 1995 | 21,847 | 11,630 | 40.0 | 10.0 | 40.0 | 10.0 | 65.0 | 10.0 | 60.0 | 10.0 |
| 1996 | 20,738 | 13,538 | 40.0 | 10.0 | 40.0 | 10.0 | 65.0 | 10.0 | 60.0 | 10.0 |
| 1997 | 21,121 | 7,756 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 1998 | 32,586 | 10,396 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 1999 | 23,904 | 6,664 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2000 | 43,151 | 14,261 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2001 | 47,339 | 19,210 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2002 | 33,087 | 14,400 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2003 | 33,371 | 20,648 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2004 | 28,506 | 15,948 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2005 | 40,628 | 14,628 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2006 | 30,979 | 21,192 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2007 | 15,735 | 18,130 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2008 | 15,696 | 16,678 | 30.0 | 10.0 | 30.0 | 10.0 | 55.0 | 10.0 | 50.0 | 10.0 |
| 2009 | 15,584 | 11,995 | 30.0 | 10.0 | 30.0 | 10.0 | 55.0 | 10.0 | 50.0 | 10.0 |
| 2010 | 22,139 | 12,175 | 30.0 | 10.0 | 30.0 | 10.0 | 50.0 | 10.0 | 40.0 | 10.0 |
| 2011 | 15,773 | 28,589 | 30.0 | 10.0 | 30.0 | 10.0 | 50.0 | 10.0 | 40.0 | 10.0 |
| 2012 | 18,582 | 23,389 | 30.0 | 10.0 | 30.0 | 10.0 | 50.0 | 10.0 | 40.0 | 10.0 |
| 2013 | 16,702 | 13,564 | 30.0 | 10.0 | 30.0 | 10.0 | 50.0 | 10.0 | 40.0 | 10.0 |
| 2014 | 15,389 | 13,699 | 30.0 | 10.0 | 30.0 | 10.0 | 40.0 | 10.0 | 35.0 | 10.0 |

Norway (Southwest)

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (SOUTH-WEST) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1972 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1973 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1974 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1975 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1976 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1977 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1978 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1979 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1980 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1981 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1982 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1983 | 31,845 | 28,601 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1984 | 23,428 | 27,641 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1985 | 29,857 | 25,515 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1986 | 29,894 | 30,769 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1987 | 30,005 | 26,623 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1988 | 36,976 | 28,255 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1989 | 19,183 | 13,041 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1990 | 18,490 | 14,423 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1991 | 9,759 | 8,323 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1992 | 6,448 | 8,832 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1993 | 11,433 | 10,239 | 40.0 | 10.0 | 40.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1994 | 18,597 | 10,961 | 40.0 | 10.0 | 40.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1995 | 10,863 | 13,122 | 40.0 | 10.0 | 40.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1996 | 7,048 | 12,546 | 40.0 | 10.0 | 40.0 | 10.0 | 70.0 | 10.0 | 65.0 | 10.0 |
| 1997 | 10,279 | 7,194 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 1998 | 5,726 | 6,583 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 1999 | 7,357 | 3,219 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2000 | 11,538 | 7,961 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2001 | 12,109 | 10,716 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2002 | 6,000 | 7,145 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2003 | 8,269 | 7,602 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2004 | 7,180 | 6,420 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2005 | 10,370 | 7,334 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2006 | 5,173 | 9,381 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2007 | 2,630 | 6,011 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2008 | 3,143 | 4,807 | 30.0 | 10.0 | 30.0 | 10.0 | 55.0 | 10.0 | 50.0 | 10.0 |
| 2009 | 3,069 | 3,792 | 30.0 | 10.0 | 30.0 | 10.0 | 55.0 | 10.0 | 50.0 | 10.0 |
| 2010 | 3,450 | 2,447 | 30.0 | 10.0 | 30.0 | 10.0 | 50.0 | 10.0 | 35.0 | 10.0 |
| 2011 | 2,888 | 4,409 | 30.0 | 10.0 | 30.0 | 10.0 | 45.0 | 10.0 | 30.0 | 10.0 |
| 2012 | 4,171 | 5,733 | 30.0 | 10.0 | 30.0 | 10.0 | 45.0 | 10.0 | 30.0 | 10.0 |
| 2013 | 3,111 | 3,581 | 30.0 | 10.0 | 30.0 | 10.0 | 45.0 | 10.0 | 30.0 | 10.0 |
| 2014 | 3,029 | 2,717 | 30.0 | 10.0 | 30.0 | 10.0 | 40.0 | 10.0 | 25.0 | 10.0 |

Mid-Norway

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (MID AREA) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1972 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1973 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1974 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1975 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1976 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1977 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1978 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1979 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1980 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1981 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1982 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1983 | 121,221 | 74,648 | 50.0 | 10.0 | 50.0 | 10.0 | 75.0 | 10.0 | 75.0 | 10.0 |
| 1984 | 94,373 | 67,639 | 50.0 | 10.0 | 50.0 | 10.0 | 75.0 | 10.0 | 75.0 | 10.0 |
| 1985 | 114,613 | 56,641 | 50.0 | 10.0 | 50.0 | 10.0 | 75.0 | 10.0 | 75.0 | 10.0 |
| 1986 | 106,921 | 77,225 | 50.0 | 10.0 | 50.0 | 10.0 | 75.0 | 10.0 | 75.0 | 10.0 |
| 1987 | 83,669 | 62,216 | 50.0 | 10.0 | 50.0 | 10.0 | 75.0 | 10.0 | 75.0 | 10.0 |
| 1988 | 80,111 | 45,609 | 50.0 | 10.0 | 50.0 | 10.0 | 75.0 | 10.0 | 75.0 | 10.0 |
| 1989 | 94,897 | 30,862 | 50.0 | 10.0 | 50.0 | 10.0 | 65.0 | 10.0 | 65.0 | 10.0 |
| 1990 | 78,888 | 40,174 | 50.0 | 10.0 | 50.0 | 10.0 | 65.0 | 10.0 | 65.0 | 10.0 |
| 1991 | 67,370 | 30,087 | 50.0 | 10.0 | 50.0 | 10.0 | 65.0 | 10.0 | 65.0 | 10.0 |
| 1992 | 51,463 | 33,092 | 50.0 | 10.0 | 50.0 | 10.0 | 65.0 | 10.0 | 65.0 | 10.0 |
| 1993 | 58,326 | 28,184 | 40.0 | 10.0 | 40.0 | 10.0 | 65.0 | 10.0 | 65.0 | 10.0 |
| 1994 | 113,427 | 33,520 | 40.0 | 10.0 | 40.0 | 10.0 | 65.0 | 10.0 | 65.0 | 10.0 |
| 1995 | 57,813 | 42,696 | 40.0 | 10.0 | 40.0 | 10.0 | 65.0 | 10.0 | 65.0 | 10.0 |
| 1996 | 28,925 | 31,613 | 40.0 | 10.0 | 40.0 | 10.0 | 65.0 | 10.0 | 65.0 | 10.0 |
| 1997 | 43,127 | 20,565 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 1998 | 63,497 | 26,817 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 1999 | 60,689 | 28,792 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2000 | 109,278 | 42,452 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2001 | 88,096 | 52,031 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2002 | 42,669 | 52,774 | 35.0 | 10.0 | 35.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2003 | 91,118 | 46,963 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2004 | 38,286 | 49,760 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2005 | 63,749 | 37,941 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2006 | 46,495 | 47,691 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2007 | 26,608 | 33,106 | 30.0 | 10.0 | 30.0 | 10.0 | 60.0 | 10.0 | 60.0 | 10.0 |
| 2008 | 31,936 | 34,869 | 30.0 | 10.0 | 30.0 | 10.0 | 55.0 | 10.0 | 45.0 | 10.0 |
| 2009 | 26,267 | 30,715 | 30.0 | 10.0 | 30.0 | 10.0 | 55.0 | 10.0 | 45.0 | 10.0 |
| 2010 | 37,557 | 30,524 | 30.0 | 10.0 | 30.0 | 10.0 | 50.0 | 10.0 | 45.0 | 10.0 |
| 2011 | 20,932 | 37,272 | 30.0 | 10.0 | 30.0 | 10.0 | 50.0 | 10.0 | 45.0 | 10.0 |
| 2012 | 22,368 | 28,265 | 30.0 | 10.0 | 30.0 | 10.0 | 50.0 | 10.0 | 45.0 | 10.0 |
| 2013 | 25,121 | 17,727 | 30.0 | 10.0 | 30.0 | 10.0 | 45.0 | 10.0 | 40.0 | 10.0 |
| 2014 | 25,349 | 14,199 | 30.0 | 10.0 | 30.0 | 10.0 | 35.0 | 10.0 | 32.0 | 10.0 |

Norway North

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (NORTH) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1972 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1973 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1974 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1975 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1976 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1977 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1978 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1979 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1980 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1981 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1982 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1983 | 104,040 | 49,413 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1984 | 150,372 | 58,858 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1985 | 118,841 | 58,956 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1986 | 84,150 | 63,418 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1987 | 72,370 | 34,232 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1988 | 53,880 | 32,140 | 50.0 | 10.0 | 50.0 | 10.0 | 80.0 | 10.0 | 80.0 | 10.0 |
| 1989 | 42,010 | 13,934 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 1990 | 38,216 | 17,321 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 1991 | 42,888 | 21,789 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 1992 | 34,593 | 19,265 | 50.0 | 10.0 | 50.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 1993 | 51,440 | 39,014 | 40.0 | 10.0 | 40.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 1994 | 37,489 | 33,411 | 40.0 | 10.0 | 40.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 1995 | 36,283 | 26,037 | 40.0 | 10.0 | 40.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 1996 | 40,792 | 36,636 | 40.0 | 10.0 | 40.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 1997 | 39,930 | 30,115 | 35.0 | 10.0 | 35.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 1998 | 46,645 | 34,806 | 35.0 | 10.0 | 35.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 1999 | 46,394 | 46,744 | 35.0 | 10.0 | 35.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 2000 | 61,854 | 51,569 | 35.0 | 10.0 | 35.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 2001 | 46,331 | 54,023 | 35.0 | 10.0 | 35.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 2002 | 38,101 | 43,100 | 35.0 | 10.0 | 35.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 2003 | 44,947 | 35,972 | 30.0 | 10.0 | 30.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 2004 | 34,640 | 28,077 | 30.0 | 10.0 | 30.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 2005 | 45,530 | 33,334 | 30.0 | 10.0 | 30.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 2006 | 48,688 | 39,508 | 30.0 | 10.0 | 30.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 2007 | 28,748 | 44,550 | 30.0 | 10.0 | 30.0 | 10.0 | 70.0 | 10.0 | 70.0 | 10.0 |
| 2008 | 34,338 | 40,553 | 30.0 | 10.0 | 30.0 | 10.0 | 65.0 | 10.0 | 65.0 | 10.0 |
| 2009 | 22,511 | 28,241 | 30.0 | 10.0 | 30.0 | 10.0 | 65.0 | 10.0 | 65.0 | 10.0 |
| 2010 | 29,836 | 28,611 | 30.0 | 10.0 | 30.0 | 10.0 | 65.0 | 10.0 | 55.0 | 10.0 |
| 2011 | 26,813 | 27,233 | 30.0 | 10.0 | 30.0 | 10.0 | 65.0 | 10.0 | 55.0 | 10.0 |
| 2012 | 28,289 | 28,000 | 30.0 | 10.0 | 30.0 | 10.0 | 65.0 | 10.0 | 55.0 | 10.0 |
| 2013 | 20,021 | 24,689 | 30.0 | 10.0 | 30.0 | 10.0 | 65.0 | 10.0 | 55.0 | 10.0 |
| 2014 | 35,171 | 23,816 | 30.0 | 10.0 | 30.0 | 10.0 | 65.0 | 10.0 | 55.0 | 10.0 |

Russia (Archangelsk and Karelia)

Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (ARCHANGEL/KORELIA). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | 134 | 16,592 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1972 | 116 | 14,434 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1973 | 169 | 20,924 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1974 | 170 | 21,137 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1975 | 140 | 17,398 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1976 | 111 | 13,781 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1977 | 78 | 9,722 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1978 | 82 | 10,134 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1979 | 112 | 13,903 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1980 | 156 | 19,397 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1981 | 68 | 8,394 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1982 | 71 | 8,797 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1983 | 48 | 11,938 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1984 | 21 | 10,680 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1985 | 454 | 11,183 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1986 | 12 | 12,291 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1987 | 647 | 8,734 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1988 | 224 | 9,978 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1989 | 989 | 10,245 | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1990 | 1,418 | 8,429 | 15.0 | 5.0 | 15.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1991 | 421 | 8,725 | 20.0 | 5.0 | 20.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1992 | 1,031 | 3,949 | 25.0 | 5.0 | 25.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1993 | 196 | 4,251 | 30.0 | 5.0 | 30.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1994 | 334 | 5,631 | 35.0 | 5.0 | 35.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1995 | 386 | 5,214 | 45.0 | 5.0 | 45.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1996 | 231 | 3,753 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1997 | 721 | 3,351 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1998 | 585 | 4,208 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 1999 | 299 | 3,101 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2000 | 514 | 3,382 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2001 | 363 | 2,348 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2002 | 1,676 | 2,439 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2003 | 893 | 2,041 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2004 | 990 | 3,761 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2005 | 1,349 | 4,915 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2006 | 2,183 | 2,841 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2007 | 1,618 | 2,621 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2008 | 332 | 2,496 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2009 | 252 | 2,214 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2010 | 397 | 3,823 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2011 | 313 | 2,585 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2012 | 1,332 | 2,446 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2013 | 2,296 | 3,480 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |
| 2014 | 2,084 | 3,463 | 55.0 | 5.0 | 55.0 | 5.0 | 60.0 | 20.0 | 60.0 | 20.0 |

Russia (Kola Peninsula: Barents Sea Basin)

Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (KOLA-BARENT SEA). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | 4,892 | 5,979 | 15.0 | 5.0 | 15.0 | 5.0 | 45.0 | 5.0 | 45.0 | 5.0 |
| 1972 | 7,978 | 9,750 | 15.0 | 5.0 | 15.0 | 5.0 | 45.0 | 5.0 | 45.0 | 5.0 |
| 1973 | 9,376 | 11,460 | 15.0 | 5.0 | 15.0 | 5.0 | 40.0 | 5.0 | 40.0 | 5.0 |
| 1974 | 12,794 | 15,638 | 15.0 | 5.0 | 15.0 | 5.0 | 40.0 | 5.0 | 40.0 | 5.0 |
| 1975 | 13,872 | 13,872 | 15.0 | 5.0 | 15.0 | 5.0 | 45.0 | 5.0 | 45.0 | 5.0 |
| 1976 | 11,493 | 14,048 | 15.0 | 5.0 | 15.0 | 5.0 | 55.0 | 5.0 | 55.0 | 5.0 |
| 1977 | 7,257 | 8,253 | 15.0 | 5.0 | 15.0 | 5.0 | 50.0 | 5.0 | 50.0 | 5.0 |
| 1978 | 7,106 | 7,113 | 15.0 | 5.0 | 15.0 | 5.0 | 55.0 | 5.0 | 55.0 | 5.0 |
| 1979 | 6,707 | 3,141 | 15.0 | 5.0 | 15.0 | 5.0 | 40.0 | 5.0 | 40.0 | 5.0 |
| 1980 | 6,621 | 5,216 | 15.0 | 5.0 | 15.0 | 5.0 | 40.0 | 5.0 | 40.0 | 5.0 |
| 1981 | 4,547 | 5,973 | 15.0 | 5.0 | 15.0 | 5.0 | 40.0 | 5.0 | 40.0 | 5.0 |
| 1982 | 5,159 | 4,798 | 15.0 | 5.0 | 15.0 | 5.0 | 35.0 | 5.0 | 35.0 | 5.0 |
| 1983 | 8,504 | 9,943 | 15.0 | 5.0 | 15.0 | 5.0 | 35.0 | 5.0 | 35.0 | 5.0 |
| 1984 | 9,453 | 12,601 | 15.0 | 5.0 | 15.0 | 5.0 | 35.0 | 5.0 | 35.0 | 5.0 |
| 1985 | 6,774 | 7,877 | 15.0 | 5.0 | 15.0 | 5.0 | 35.0 | 5.0 | 35.0 | 5.0 |
| 1986 | 10,147 | 5,352 | 15.0 | 5.0 | 15.0 | 5.0 | 40.0 | 5.0 | 40.0 | 5.0 |
| 1987 | 8,560 | 5,149 | 15.0 | 5.0 | 15.0 | 5.0 | 40.0 | 5.0 | 40.0 | 5.0 |
| 1988 | 6,644 | 3,655 | 15.0 | 5.0 | 15.0 | 5.0 | 35.0 | 5.0 | 35.0 | 5.0 |
| 1989 | 13,424 | 6,787 | 15.0 | 5.0 | 15.0 | 5.0 | 40.0 | 5.0 | 40.0 | 5.0 |
| 1990 | 16,038 | 8,234 | 15.0 | 5.0 | 15.0 | 5.0 | 40.0 | 5.0 | 40.0 | 5.0 |
| 1991 | 4,550 | 7,568 | 15.0 | 5.0 | 15.0 | 5.0 | 30.0 | 5.0 | 30.0 | 5.0 |
| 1992 | 11,394 | 7,109 | 15.0 | 5.0 | 15.0 | 5.0 | 30.0 | 5.0 | 30.0 | 5.0 |
| 1993 | 8,642 | 5,690 | 15.0 | 5.0 | 15.0 | 5.0 | 30.0 | 5.0 | 30.0 | 5.0 |
| 1994 | 6,101 | 4,632 | 15.0 | 5.0 | 15.0 | 5.0 | 30.0 | 5.0 | 30.0 | 5.0 |
| 1995 | 6,318 | 3,693 | 15.0 | 5.0 | 15.0 | 5.0 | 30.0 | 5.0 | 30.0 | 5.0 |
| 1996 | 6,815 | 1,701 | 20.0 | 5.0 | 20.0 | 5.0 | 25.0 | 5.0 | 25.0 | 5.0 |
| 1997 | 3,564 | 867 | 25.0 | 5.0 | 25.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 1998 | 1,854 | 280 | 35.0 | 5.0 | 35.0 | 5.0 | 12.5 | 2.5 | 12.5 | 2.5 |
| 1999 | 1,510 | 424 | 40.0 | 5.0 | 40.0 | 5.0 | 7.5 | 2.5 | 7.5 | 2.5 |
| 2000 | 805 | 323 | 50.0 | 5.0 | 50.0 | 5.0 | 6.0 | 2.0 | 6.0 | 2.0 |
| 2001 | 591 | 241 | 60.0 | 5.0 | 60.0 | 5.0 | 3.5 | 1.5 | 3.5 | 1.5 |
| 2002 | 1,436 | 2,478 | 50.0 | 10.0 | 50.0 | 10.0 | 10.0 | 5.0 | 20.0 | 5.0 |
| 2003 | 1,938 | 1,095 | 50.0 | 10.0 | 50.0 | 10.0 | 10.0 | 5.0 | 20.0 | 5.0 |
| 2004 | 1,095 | 850 | 50.0 | 10.0 | 50.0 | 10.0 | 10.0 | 5.0 | 20.0 | 5.0 |
| 2005 | 859 | 426 | 60.0 | 10.0 | 60.0 | 10.0 | 10.0 | 5.0 | 20.0 | 5.0 |
| 2006 | 1,372 | 844 | 60.0 | 10.0 | 60.0 | 10.0 | 10.0 | 5.0 | 20.0 | 5.0 |
| 2007 | 784 | 707 | 60.0 | 10.0 | 60.0 | 10.0 | 10.0 | 5.0 | 20.0 | 5.0 |
| 2008 | 1,446 | 997 | 60.0 | 10.0 | 60.0 | 10.0 | 15.0 | 5.0 | 20.0 | 5.0 |
| 2009 | 2,882 | 1,080 | 60.0 | 10.0 | 60.0 | 10.0 | 15.0 | 5.0 | 20.0 | 5.0 |
| 2010 | 3,884 | 1,486 | 60.0 | 10.0 | 60.0 | 10.0 | 20.0 | 5.0 | 25.0 | 5.0 |
| 2011 | 3,861 | 1,407 | 60.0 | 10.0 | 60.0 | 10.0 | 20.0 | 5.0 | 25.0 | 5.0 |
| 2012 | 2,708 | 1,027 | 60.0 | 10.0 | 60.0 | 10.0 | 20.0 | 5.0 | 25.0 | 5.0 |
| 2013 | 939 | 904 | 60.0 | 10.0 | 60.0 | 10.0 | 20.0 | 5.0 | 25.0 | 5.0 |
| 2014 | 969 | 789 | 60.0 | 10.0 | 60.0 | 10.0 | 20.0 | 5.0 | 25.0 | 5.0 |

Russia (Kola Peninsula: White Sea Basin)

| Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (KOLA-WHITE SEA). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation) | | | | | | | | | | | | |
|----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|---------------------------|---------------------------|-----------------------------------|-----------------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Catch 1SW following-year spawners | Catch MSW following-year spawners | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
| 1971 | 67,845 | 29,077 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1972 | 45,837 | 19,644 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1973 | 68,684 | 29,436 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1974 | 63,892 | 27,382 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1975 | 109,038 | 46,730 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1976 | 76,281 | 41,075 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1977 | 47,943 | 32,392 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1978 | 49,291 | 17,307 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1979 | 69,511 | 21,369 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1980 | 46,037 | 23,241 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1981 | 40,172 | 12,747 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1982 | 32,619 | 14,840 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1983 | 54,217 | 20,840 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1984 | 56,786 | 16,893 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1985 | 87,274 | 16,876 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1986 | 72,102 | 17,681 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 60.0 | 10.0 |
| 1987 | 79,639 | 12,501 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 1988 | 44,813 | 18,777 | 0.0 | 0.0 | 3.0 | 2.0 | 3.0 | 2.0 | 45.0 | 5.0 | 45.0 | 5.0 |
| 1989 | 53,293 | 11,448 | 0.0 | 0.0 | 7.5 | 2.5 | 7.5 | 2.5 | 45.0 | 5.0 | 45.0 | 5.0 |
| 1990 | 44,409 | 11,152 | 0.0 | 0.0 | 12.5 | 2.5 | 12.5 | 2.5 | 45.0 | 5.0 | 45.0 | 5.0 |
| 1991 | 31,978 | 6,263 | 0.0 | 0.0 | 17.5 | 2.5 | 17.5 | 2.5 | 35.0 | 5.0 | 35.0 | 5.0 |
| 1992 | 23,827 | 3,680 | 0.0 | 0.0 | 22.5 | 2.5 | 22.5 | 2.5 | 25.0 | 5.0 | 25.0 | 5.0 |
| 1993 | 20,987 | 5,552 | 0.0 | 0.0 | 25.0 | 5.0 | 25.0 | 5.0 | 25.0 | 5.0 | 25.0 | 5.0 |
| 1994 | 25,178 | 3,680 | 0.0 | 0.0 | 30.0 | 5.0 | 30.0 | 5.0 | 25.0 | 5.0 | 15.0 | 5.0 |
| 1995 | 19,381 | 2,847 | 0.0 | 0.0 | 35.0 | 5.0 | 35.0 | 5.0 | 25.0 | 5.0 | 15.0 | 5.0 |
| 1996 | 27,097 | 2,710 | 0.0 | 0.0 | 35.0 | 5.0 | 35.0 | 5.0 | 25.0 | 5.0 | 15.0 | 5.0 |
| 1997 | 27,695 | 2,085 | 0.0 | 0.0 | 35.0 | 5.0 | 35.0 | 5.0 | 25.0 | 5.0 | 15.0 | 5.0 |
| 1998 | 32,693 | 1,963 | 0.0 | 0.0 | 35.0 | 5.0 | 35.0 | 5.0 | 25.0 | 5.0 | 15.0 | 5.0 |
| 1999 | 22,330 | 2,841 | 0.0 | 0.0 | 35.0 | 5.0 | 35.0 | 5.0 | 25.0 | 5.0 | 15.0 | 5.0 |
| 2000 | 26,376 | 4,396 | 0.0 | 0.0 | 35.0 | 5.0 | 35.0 | 5.0 | 25.0 | 5.0 | 15.0 | 5.0 |
| 2001 | 20,483 | 3,959 | 0.0 | 0.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2002 | 19,174 | 3,937 | 0.0 | 0.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2003 | 15,687 | 3,734 | 0.0 | 0.0 | 35.0 | 5.0 | 25.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2004 | 10,947 | 1,990 | 0.0 | 0.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2005 | 13,172 | 2,388 | 1212.0 | 878.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2006 | 15,004 | 2,071 | 3852.0 | 399.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2007 | 7,807 | 1,404 | 2264.0 | 852.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2008 | 8,447 | 4,711 | 3175.0 | 832.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2009 | 5,351 | 3,105 | 5130.0 | 1710.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2010 | 6,731 | 4,158 | 3684.0 | 1228.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2011 | 7,363 | 4,325 | 3082.0 | 1027.3 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2012 | 10,398 | 1,431 | 2267.0 | 756.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2013 | 8,986 | 1,660 | 2203.0 | 734.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2014 | 8,593 | 1,674 | 3307.0 | 1102.0 | 35.0 | 5.0 | 35.0 | 5.0 | 15.0 | 5.0 | 15.0 | 5.0 |

Russia (Pechora River)

| Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (PECHORA RIVER). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation) | | | | | | | | | | | | | |
|---------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|---------------------------|---------------------------|----------------------------|----------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|--|
| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Return estimate 1SW salmon | Return estimate MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon | |
| 1971 | 605 | 17,728 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1972 | 825 | 24,175 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1973 | 1,705 | 49,962 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1974 | 1,320 | 38,680 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1975 | 1,298 | 38,046 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1976 | 991 | 34,394 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1977 | 589 | 20,464 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1978 | 759 | 26,341 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1979 | 421 | 14,614 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1980 | 1,123 | 39,001 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1981 | 126 | 20,874 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1982 | 54 | 13,546 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1983 | 598 | 16,002 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1984 | 1,833 | 15,967 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1985 | 2,763 | 29,738 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1986 | 66 | 32,734 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1987 | 21 | 21,179 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1988 | 3,184 | 12,816 | NA | NA | 20.0 | 10.0 | 20.0 | 10.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1989 | NA | NA | 24,596 | 27,404 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1990 | NA | NA | 50 | 49,950 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1991 | NA | NA | 7,975 | 47,025 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1992 | NA | NA | 550 | 54,450 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1993 | NA | NA | 68 | 67,932 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1994 | NA | NA | 3,900 | 48,100 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1995 | NA | NA | 9,280 | 70,720 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1996 | NA | NA | 8,664 | 48,336 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1997 | NA | NA | 1,440 | 38,560 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1998 | NA | NA | 780 | 59,220 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 1999 | NA | NA | 2,120 | 37,880 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2000 | NA | NA | 84 | 83,916 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2001 | NA | NA | 2,244 | 41,756 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2002 | NA | NA | 405 | 44,595 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2003 | NA | NA | 1,650 | 31,350 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2004 | NA | NA | 6,075 | 20,925 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2005 | NA | NA | 2,852 | 28,148 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2006 | NA | NA | 1,472 | 30,528 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2007 | NA | NA | 817 | 42,183 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2008 | NA | NA | 300 | 49,700 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2009 | NA | NA | 1,116 | 47,385 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2010 | NA | NA | 1,096 | 53,704 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2011 | NA | NA | 2,990 | 56,810 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2012 | NA | NA | 4,424 | 27,176 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2013 | NA | NA | 4,225 | 30,983 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |
| 2014 | NA | NA | 2,251 | 31,349 | 10.0 | 5.0 | 10.0 | 5.0 | 65.0 | 15.0 | 65.0 | 15.0 | |

Sweden

Annual input data for NEAC PFA run-reconstruction & NCL models for SWEDEN. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | 6,220 | 254 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1972 | 4,943 | 201 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1973 | 6,124 | 895 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1974 | 8,870 | 563 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1975 | 9,620 | 160 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1976 | 5,420 | 480 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1977 | 2,453 | 206 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1978 | 2,903 | 254 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1979 | 2,988 | 661 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1980 | 3,842 | 1,283 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1981 | 7,013 | 284 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1982 | 6,177 | 1,381 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1983 | 8,222 | 903 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1984 | 11,584 | 1,266 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1985 | 13,810 | 470 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1986 | 14,415 | 240 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1987 | 11,450 | 1,084 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1988 | 9,604 | 1,160 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1989 | 2,803 | 4,044 | 30.0 | 15.0 | 30.0 | 15.0 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1990 | 6,839 | 2,249 | 15.0 | 10.0 | 15.0 | 10.0 | 45.0 | 15.0 | 50.0 | 15.0 |
| 1991 | 8,599 | 3,033 | 15.0 | 10.0 | 15.0 | 10.0 | 45.0 | 15.0 | 50.0 | 15.0 |
| 1992 | 9,550 | 4,205 | 15.0 | 10.0 | 15.0 | 10.0 | 45.0 | 15.0 | 50.0 | 15.0 |
| 1993 | 9,468 | 4,762 | 15.0 | 10.0 | 15.0 | 10.0 | 45.0 | 15.0 | 50.0 | 15.0 |
| 1994 | 7,347 | 3,628 | 15.0 | 10.0 | 15.0 | 10.0 | 45.0 | 15.0 | 50.0 | 15.0 |
| 1995 | 8,933 | 1,528 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 1996 | 5,318 | 2,507 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 1997 | 2,415 | 1,809 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 1998 | 1,953 | 1,000 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 1999 | 3,075 | 712 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2000 | 5,660 | 2,546 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2001 | 3,504 | 3,026 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2002 | 3,374 | 2,075 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2003 | 1,833 | 496 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2004 | 1,537 | 1,528 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2005 | 1,503 | 1,027 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2006 | 1,676 | 1,069 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2007 | 521 | 1,001 | 15.0 | 10.0 | 15.0 | 10.0 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2008 | 615 | 1,112 | 12.5 | 7.5 | 12.5 | 7.5 | 27.5 | 12.5 | 32.5 | 12.5 |
| 2009 | 651 | 979 | 12.5 | 7.5 | 12.5 | 7.5 | 27.5 | 12.5 | 32.5 | 12.5 |
| 2010 | 1,111 | 1,139 | 12.5 | 7.5 | 12.5 | 7.5 | 27.5 | 12.5 | 32.5 | 12.5 |
| 2011 | 1,460 | 3,100 | 17.5 | 7.5 | 17.5 | 7.5 | 45.0 | 15.0 | 50.0 | 15.0 |
| 2012 | 1,336 | 3,130 | 12.5 | 7.5 | 10.0 | 5.0 | 27.5 | 12.5 | 32.5 | 12.5 |
| 2013 | 874 | 1,431 | 10.0 | 5.0 | 10.0 | 5.0 | 30.0 | 15.0 | 35.0 | 15.0 |
| 2014 | 2,347 | 2,797 | 12.5 | 7.5 | 12.5 | 7.5 | 30.0 | 12.5 | 35.0 | 12.5 |

UK (England and Wales)

| Annual input data for NEAC PFA run-reconstruction & NCL models for UK(ENGLAND AND WALES). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation) | | | | | | | | | | | | | | | | | | |
|--------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|----------------------|----------------------------------|--------------------------------------------|---------------------------------------------------|-------------------------------------------------|---------------------------------------|-------------------------------------|--------------------------------------|-------------------------------------|--------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|--------------------------------------------|--------------------------------------------|--------------------------------------------------|--------------------------------------------------|--------------------------------------------------------|
| Year | Declared total catch | Estimated proportion 1SW (total) | Declared catch in NE coastal fishery total | Declared catch in NE coastal fishery - drift nets | Declared catch in NE coastal fishery - T/J nets | Estimated proportion 1SW (NE fishery) | Estimated % unreported catch of 1SW | Uncertainty in % unreported catch of | Estimated % unreported catch of MSW | Uncertainty in % unreported catch of | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW | Estimated % unreported catch in NE fishery | Estimated proportion Scottish fish in NE fishery | Estimated proportion Scottish fish in NE fishery | Estimated proportion Scottish fish in NE fishery (T/J) |
| 1971 | 109,861 | 0.55 | 60,353 | NA | NA | 0.55 | 38.3 | 9.6 | 38.3 | 9.6 | 57.3 | 10.0 | 42.5 | 10.0 | 32.3 | 0.95 | NA | NA |
| 1972 | 108,074 | 0.42 | 51,681 | NA | NA | 0.42 | 39.0 | 9.7 | 39.0 | 9.7 | 51.3 | 10.0 | 37.8 | 10.0 | 32.3 | 0.95 | NA | NA |
| 1973 | 114,786 | 0.53 | 62,842 | NA | NA | 0.53 | 38.4 | 9.6 | 38.4 | 9.6 | 50.6 | 10.0 | 37.3 | 10.0 | 32.3 | 0.95 | NA | NA |
| 1974 | 104,325 | 0.65 | 52,756 | NA | NA | 0.65 | 39.3 | 9.8 | 39.3 | 9.8 | 50.2 | 10.0 | 37.0 | 10.0 | 32.3 | 0.95 | NA | NA |
| 1975 | 113,062 | 0.59 | 53,451 | NA | NA | 0.59 | 38.5 | 9.6 | 38.5 | 9.6 | 49.8 | 10.0 | 36.7 | 10.0 | 32.3 | 0.95 | NA | NA |
| 1976 | 54,294 | 0.64 | 15,701 | NA | NA | 0.64 | 36.8 | 9.2 | 36.8 | 9.2 | 50.3 | 10.0 | 37.1 | 10.0 | 32.3 | 0.94 | NA | NA |
| 1977 | 94,282 | 0.62 | 52,888 | NA | NA | 0.62 | 39.0 | 9.8 | 39.0 | 9.8 | 50.4 | 10.0 | 37.2 | 10.0 | 32.3 | 0.93 | NA | NA |
| 1978 | 93,125 | 0.69 | 51,630 | NA | NA | 0.69 | 38.4 | 9.6 | 38.4 | 9.6 | 49.1 | 10.0 | 36.2 | 10.0 | 32.3 | 0.92 | NA | NA |
| 1979 | 75,386 | 0.81 | 43,464 | NA | NA | 0.81 | 38.6 | 9.6 | 38.6 | 9.6 | 47.7 | 10.0 | 35.2 | 10.0 | 32.3 | 0.91 | NA | NA |
| 1980 | 90,218 | 0.55 | 45,780 | NA | NA | 0.55 | 39.1 | 9.8 | 39.1 | 9.8 | 47.8 | 10.0 | 35.2 | 10.0 | 32.3 | 0.90 | NA | NA |
| 1981 | 121,039 | 0.48 | 69,113 | NA | NA | 0.48 | 38.3 | 9.6 | 38.3 | 9.6 | 47.4 | 10.0 | 34.9 | 10.0 | 32.3 | 0.89 | NA | NA |
| 1982 | 80,289 | 0.67 | 50,167 | NA | NA | 0.67 | 38.3 | 9.6 | 38.3 | 9.6 | 47.3 | 10.0 | 34.8 | 10.0 | 32.3 | 0.88 | NA | NA |
| 1983 | 116,995 | 0.72 | 77,277 | NA | NA | 0.72 | 37.1 | 9.3 | 37.1 | 9.3 | 47.1 | 10.0 | 34.7 | 10.0 | 32.3 | 0.87 | NA | NA |
| 1984 | 94,271 | 0.74 | 59,295 | NA | NA | 0.74 | 36.5 | 9.1 | 36.5 | 9.1 | 47.4 | 10.0 | 34.8 | 10.0 | 32.3 | 0.86 | NA | NA |
| 1985 | 95,531 | 0.66 | 57,356 | NA | NA | 0.66 | 38.9 | 9.7 | 38.9 | 9.7 | 47.5 | 10.0 | 34.9 | 10.0 | 32.3 | 0.85 | NA | NA |
| 1986 | 110,794 | 0.62 | 63,425 | NA | NA | 0.62 | 38.0 | 9.5 | 38.0 | 9.5 | 46.9 | 10.0 | 34.3 | 10.0 | 32.3 | 0.84 | NA | NA |
| 1987 | 83,439 | 0.68 | 36,143 | NA | NA | 0.68 | 38.2 | 9.5 | 38.2 | 9.5 | 46.1 | 10.0 | 33.7 | 10.0 | 32.3 | 0.83 | NA | NA |
| 1988 | 110,163 | 0.69 | NA | 47,465 | 3,384 | 0.69 | 39.7 | 9.9 | 39.7 | 9.9 | 45.5 | 10.0 | 33.5 | 10.0 | 32.3 | NA | 0.82 | 0.50 |
| 1989 | 83,668 | 0.65 | NA | 36,236 | 5,217 | 0.65 | 36.9 | 9.2 | 36.9 | 9.2 | 45.3 | 10.0 | 33.3 | 10.0 | 32.3 | NA | 0.81 | 0.50 |
| 1990 | 86,676 | 0.52 | NA | 48,219 | 3,311 | 0.52 | 36.7 | 9.2 | 36.7 | 9.2 | 45.3 | 10.0 | 33.2 | 10.0 | 31.3 | NA | 0.80 | 0.50 |
| 1991 | 51,649 | 0.71 | NA | 22,463 | 2,966 | 0.71 | 37.3 | 9.3 | 37.3 | 9.3 | 44.0 | 10.0 | 32.3 | 10.0 | 29.7 | NA | 0.79 | 0.50 |
| 1992 | 44,586 | 0.77 | NA | 17,574 | 2,570 | 0.77 | 39.8 | 10.0 | 39.8 | 10.0 | 43.5 | 10.0 | 31.8 | 10.0 | 28.0 | NA | 0.78 | 0.50 |
| 1993 | 69,177 | 0.81 | NA | 39,224 | 2,576 | 0.81 | 38.0 | 9.5 | 38.0 | 9.5 | 40.6 | 10.0 | 29.5 | 10.0 | 26.3 | NA | 0.77 | 0.50 |
| 1994 | 88,121 | 0.77 | NA | 41,298 | 5,256 | 0.77 | 23.9 | 6.0 | 23.9 | 6.0 | 40.5 | 10.0 | 29.5 | 10.0 | 24.4 | NA | 0.76 | 0.50 |
| 1995 | 80,478 | 0.72 | NA | 48,005 | 5,205 | 0.72 | 22.3 | 5.6 | 22.3 | 5.6 | 37.6 | 10.0 | 27.1 | 10.0 | 22.5 | NA | 0.75 | 0.50 |
| 1996 | 46,696 | 0.65 | NA | 15,172 | 3,409 | 0.65 | 20.6 | 5.1 | 20.6 | 5.1 | 35.8 | 10.0 | 25.8 | 10.0 | 20.6 | NA | 0.75 | 0.50 |
| 1997 | 41,374 | 0.73 | NA | 19,241 | 2,681 | 0.73 | 18.8 | 4.7 | 18.8 | 4.7 | 33.4 | 10.0 | 23.9 | 10.0 | 18.5 | NA | 0.75 | 0.50 |
| 1998 | 36,917 | 0.82 | NA | 17,328 | 937 | 0.82 | 18.9 | 4.7 | 18.9 | 4.7 | 31.4 | 10.0 | 22.4 | 10.0 | 18.5 | NA | 0.75 | 0.50 |
| 1999 | 41,094 | 0.68 | NA | 24,812 | 2,021 | 0.68 | 17.4 | 4.4 | 17.4 | 4.4 | 29.5 | 10.0 | 17.9 | 9.0 | 17.1 | NA | 0.75 | 0.50 |
| 2000 | 60,953 | 0.79 | NA | 40,059 | 3,295 | 0.79 | 14.9 | 3.7 | 14.9 | 3.7 | 29.7 | 10.0 | 15.0 | 7.5 | 13.1 | NA | 0.75 | 0.50 |
| 2001 | 51,307 | 0.75 | NA | 32,374 | 3,741 | 0.75 | 14.8 | 3.7 | 14.8 | 3.7 | 27.9 | 10.0 | 14.3 | 7.1 | 13.1 | NA | 0.75 | 0.50 |
| 2002 | 45,669 | 0.76 | NA | 27,685 | 3,295 | 0.76 | 15.3 | 3.8 | 15.3 | 3.8 | 27.8 | 10.0 | 14.1 | 7.0 | 13.9 | NA | 0.75 | 0.50 |
| 2003 | 22,206 | 0.66 | NA | 5,511 | 4,924 | 0.66 | 17.4 | 4.4 | 17.4 | 4.4 | 21.4 | 10.0 | 10.7 | 5.3 | 17.1 | NA | 0.75 | 0.50 |
| 2004 | 30,559 | 0.81 | NA | 5,921 | 5,096 | 0.81 | 17.7 | 4.4 | 17.7 | 4.4 | 22.1 | 10.0 | 10.6 | 5.3 | 17.1 | NA | 0.75 | 0.50 |
| 2005 | 26,162 | 0.76 | NA | 5,607 | 3,380 | 0.76 | 17.6 | 4.4 | 17.6 | 4.4 | 21.8 | 10.0 | 10.6 | 5.3 | 17.1 | NA | 0.75 | 0.50 |
| 2006 | 22,056 | 0.78 | NA | 4,040 | 3,526 | 0.78 | 17.6 | 4.4 | 17.6 | 4.4 | 19.5 | 9.8 | 9.1 | 4.6 | 17.1 | NA | 0.75 | 0.50 |
| 2007 | 19,914 | 0.78 | NA | 4,894 | 2,197 | 0.78 | 17.7 | 4.4 | 17.7 | 4.4 | 17.9 | 9.0 | 8.4 | 4.2 | 17.1 | NA | 0.75 | 0.50 |
| 2008 | 19,036 | 0.76 | NA | 3,649 | 2,592 | 0.76 | 17.8 | 4.4 | 17.8 | 4.4 | 17.6 | 8.8 | 8.2 | 4.1 | 17.1 | NA | 0.75 | 0.50 |
| 2009 | 13,910 | 0.72 | NA | 2,590 | 2,805 | 0.72 | 11.4 | 2.9 | 11.4 | 2.9 | 17.4 | 8.7 | 8.2 | 4.1 | 7.4 | NA | 0.75 | 0.50 |
| 2010 | 32,695 | 0.78 | NA | 12,214 | 7,768 | 0.78 | 10.8 | 2.7 | 10.8 | 2.7 | 17.5 | 8.8 | 8.0 | 4.0 | 7.4 | NA | 0.75 | 0.50 |
| 2011 | 34,575 | 0.57 | NA | 14,915 | 9,233 | 0.57 | 10.5 | 2.6 | 10.5 | 2.6 | 20.8 | 10.0 | 10.2 | 5.1 | 7.4 | NA | 0.75 | 0.50 |
| 2012 | 14,926 | 0.50 | NA | 3,571 | 3,705 | 0.50 | 11.5 | 2.9 | 11.5 | 2.9 | 16.8 | 8.4 | 8.0 | 4.0 | 7.4 | NA | 0.75 | 0.50 |
| 2013 | 22,608 | 0.58 | NA | 7,964 | 8,679 | 0.58 | 9.9 | 2.5 | 9.9 | 2.5 | 17.4 | 8.7 | 8.5 | 4.3 | 7.4 | NA | 0.75 | 0.50 |
| 2014 | 13,502 | 0.54 | NA | 6,753 | 3,457 | 0.54 | 9.6 | 2.4 | 9.6 | 2.4 | 16.0 | 8.0 | 8.1 | 4.0 | 7.4 | NA | 0.75 | 0.50 |

UK (Northern Ireland)-Foyle Fisheries Area

Annual input data for NEAC PFA run-reconstruction & NCL models for UK(NORTHERN IRELAND) (FOYLE). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared net catch 1SW salmon | Declared net catch MSW salmon | Declared rod catch 1SW salmon | Declared rod catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|-------------------------------|-------------------------------|-------------------------------|-------------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | 78,037 | 5,874 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1972 | 64,663 | 4,867 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1973 | 57,469 | 4,326 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1974 | 72,587 | 5,464 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1975 | 51,061 | 3,843 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1976 | 36,206 | 2,725 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1977 | 36,510 | 2,748 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1978 | 44,557 | 3,354 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1979 | 34,413 | 2,590 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1980 | 45,777 | 3,446 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1981 | 32,346 | 2,435 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1982 | 55,946 | 4,211 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1983 | 77,424 | 5,828 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1984 | 27,465 | 2,067 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1985 | 37,685 | 2,836 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1986 | 43,109 | 3,245 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1987 | 17,189 | 1,294 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 69.0 | 7.0 | 46.0 | 5.0 |
| 1988 | 43,974 | 3,310 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 64.5 | 6.5 | 36.0 | 4.0 |
| 1989 | 60,288 | 4,538 | NA | NA | 23.5 | 13.5 | 23.5 | 13.5 | 89.0 | 9.0 | 60.0 | 6.0 |
| 1990 | 39,875 | 3,001 | NA | NA | 13.5 | 3.5 | 13.5 | 3.5 | 62.0 | 6.0 | 38.0 | 4.0 |
| 1991 | 21,709 | 1,634 | NA | NA | 13.5 | 3.5 | 13.5 | 3.5 | 64.5 | 6.5 | 43.0 | 4.0 |
| 1992 | 39,299 | 2,958 | NA | NA | 16.5 | 6.5 | 16.5 | 6.5 | 56.0 | 6.0 | 33.0 | 3.0 |
| 1993 | 35,366 | 2,662 | NA | NA | 13.5 | 3.5 | 13.5 | 3.5 | 41.0 | 4.0 | 12.0 | 1.0 |
| 1994 | 36,144 | 2,720 | NA | NA | 19.0 | 9.0 | 19.0 | 9.0 | 70.0 | 7.0 | 40.0 | 4.0 |
| 1995 | 33,398 | 2,514 | NA | NA | 13.5 | 3.5 | 13.5 | 3.5 | 67.0 | 7.0 | 42.0 | 4.0 |
| 1996 | 28,406 | 2,138 | NA | NA | 15.0 | 5.0 | 15.0 | 5.0 | 57.0 | 10.0 | 34.0 | 10.0 |
| 1997 | 40,886 | 3,077 | NA | NA | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 10.0 | 34.0 | 10.0 |
| 1998 | 37,154 | 2,797 | NA | NA | 10.0 | 5.0 | 10.0 | 5.0 | 25.0 | 5.0 | 22.5 | 7.5 |
| 1999 | 21,660 | 1,630 | NA | NA | 10.0 | 5.0 | 10.0 | 5.0 | 63.0 | 5.0 | 32.5 | 7.5 |
| 2000 | 30,385 | 2,287 | NA | NA | 10.0 | 5.0 | 10.0 | 5.0 | 58.0 | 5.0 | 32.5 | 7.5 |
| 2001 | 21,368 | 1,608 | NA | NA | 5.0 | 5.0 | 5.0 | 5.0 | 50.0 | 5.0 | 30.0 | 5.0 |
| 2002 | 37,914 | 2,854 | 9163.0 | 690.0 | 2.5 | 2.5 | 2.5 | 2.5 | 15.0 | 3.0 | 15.0 | 3.0 |
| 2003 | 30,441 | 2,291 | 4576.0 | 344.0 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 | 15.0 | 3.0 |
| 2004 | 20,730 | 1,560 | 4570.0 | 344.0 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 | 15.0 | 3.0 |
| 2005 | 23,746 | 1,787 | 7079.0 | 533.0 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 | 15.0 | 3.0 |
| 2006 | 11,324 | 852 | 4886.0 | 368.0 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 | 15.0 | 3.0 |
| 2007 | 5,050 | 322 | 9530.0 | 608.0 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 | 15.0 | 3.0 |
| 2008 | 3,880 | 292 | 4755.0 | 304.0 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 | 15.0 | 3.0 |
| 2009 | 1,743 | 194 | 3640.0 | 405.0 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 | 15.0 | 3.0 |
| 2010 | - | - | 4257.0 | 473.0 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 | 15.0 | 3.0 |
| 2011 | - | - | 3770.0 | 1256.0 | 1.0 | 1.0 | 1.0 | 1.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2012 | - | - | 4781.0 | 1594.0 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 7.5 | 10.0 | 7.5 |
| 2013 | - | - | 2831.0 | 283.0 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 7.5 | 10.0 | 7.5 |
| 2014 | - | - | 3036.0 | 856.0 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 7.5 | 10.0 | 7.5 |

UK (Northern Ireland)-DCAL area

Annual input data for NEAC PFA run-reconstruction & NCL models for UK(NORTHERN IRELAND) (LOCHS AGENCY AREA) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared net catch 1SW salmon | Declared net catch MSW salmon | Declared rod catch 1SW salmon | Declared rod catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|-------------------------------|-------------------------------|-------------------------------|-------------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | 35,506 | 2,673 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1972 | 34,550 | 2,601 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1973 | 29,229 | 2,200 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1974 | 22,307 | 1,679 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1975 | 26,701 | 2,010 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1976 | 17,886 | 1,346 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1977 | 16,778 | 1,263 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1978 | 24,857 | 1,871 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1979 | 14,323 | 1,078 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1980 | 15,967 | 1,202 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1981 | 15,994 | 1,204 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1982 | 14,068 | 1,059 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1983 | 20,845 | 1,569 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1984 | 11,109 | 836 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1985 | 12,369 | 931 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1986 | 13,160 | 991 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 | 50.0 | 5.0 |
| 1987 | 9,240 | 695 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 69.0 | 7.0 | 46.0 | 5.0 |
| 1988 | 14,320 | 1,078 | NA | NA | 21.5 | 11.5 | 21.5 | 11.5 | 64.5 | 6.5 | 36.0 | 4.0 |
| 1989 | 15,081 | 1,135 | NA | NA | 23.5 | 13.5 | 23.5 | 13.5 | 89.0 | 9.0 | 60.0 | 6.0 |
| 1990 | 9,499 | 715 | NA | NA | 13.5 | 3.5 | 13.5 | 3.5 | 62.0 | 6.0 | 38.0 | 4.0 |
| 1991 | 6,987 | 526 | NA | NA | 13.5 | 3.5 | 13.5 | 3.5 | 64.5 | 6.5 | 43.0 | 4.0 |
| 1992 | 9,346 | 703 | NA | NA | 16.5 | 6.5 | 16.5 | 6.5 | 56.0 | 6.0 | 33.0 | 3.0 |
| 1993 | 7,906 | 595 | NA | NA | 13.5 | 3.5 | 13.5 | 3.5 | 41.0 | 4.0 | 12.0 | 1.0 |
| 1994 | 11,206 | 843 | NA | NA | 19.0 | 9.0 | 19.0 | 9.0 | 70.0 | 7.0 | 40.0 | 4.0 |
| 1995 | 11,637 | 876 | NA | NA | 13.5 | 3.5 | 13.5 | 3.5 | 67.0 | 7.0 | 42.0 | 4.0 |
| 1996 | 10,383 | 781 | NA | NA | 15.0 | 5.0 | 15.0 | 5.0 | 57.0 | 10.0 | 34.0 | 10.0 |
| 1997 | 10,479 | 789 | NA | NA | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 10.0 | 34.0 | 10.0 |
| 1998 | 9,375 | 706 | NA | NA | 10.0 | 5.0 | 10.0 | 5.0 | 25.0 | 5.0 | 22.5 | 7.5 |
| 1999 | 9,011 | 678 | NA | NA | 10.0 | 5.0 | 10.0 | 5.0 | 63.0 | 5.0 | 32.5 | 7.5 |
| 2000 | 10,598 | 798 | NA | NA | 10.0 | 5.0 | 10.0 | 5.0 | 58.0 | 5.0 | 32.5 | 7.5 |
| 2001 | 8,104 | 610 | NA | NA | 5.0 | 5.0 | 5.0 | 5.0 | 50.0 | 5.0 | 30.0 | 5.0 |
| 2002 | 3,315 | 249 | 2,218 | 167 | 2.5 | 2.5 | 2.5 | 2.5 | 13.7 | 8.8 | 13.7 | 8.8 |
| 2003 | 2,236 | 168 | 1,884 | 141 | 2.5 | 2.5 | 2.5 | 2.5 | 12.3 | 6.6 | 12.3 | 6.6 |
| 2004 | 2,411 | 181 | 3,053 | 230 | 0.5 | 0.5 | 0.5 | 0.5 | 18.3 | 9.7 | 18.3 | 9.7 |
| 2005 | 3,012 | 227 | 1,791 | 135 | 0.5 | 0.5 | 0.5 | 0.5 | 11.9 | 7.1 | 11.9 | 7.1 |
| 2006 | 2,288 | 172 | 1,289 | 97 | 0.5 | 0.5 | 0.5 | 0.5 | 12.4 | 8.0 | 12.4 | 8.0 |
| 2007 | 2,533 | 162 | 2,427 | 155 | 0.5 | 0.5 | 0.5 | 0.5 | 11.0 | 3.6 | 11.0 | 3.6 |
| 2008 | 1,825 | 116 | 2,444 | 156 | 0.5 | 0.5 | 0.5 | 0.5 | 13.9 | 7.1 | 13.9 | 7.1 |
| 2009 | 1,383 | 154 | 1,457 | 162 | 0.5 | 0.5 | 0.5 | 0.5 | 9.9 | 3.0 | 9.9 | 3.0 |
| 2010 | 1,723 | 191 | 1,327 | 147 | 0.5 | 0.5 | 0.5 | 0.5 | 14.6 | 2.5 | 14.6 | 2.5 |
| 2011 | 857 | 285 | 1,132 | 378 | 1.0 | 1.0 | 1.0 | 1.0 | 15.0 | 5.0 | 15.0 | 5.0 |
| 2012 | 15 | 5 | 263 | 87 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 5.0 | 10.0 | 5.0 |
| 2013 | 9 | 1 | 46 | 5 | 1.0 | 1.0 | 1.0 | 1.0 | 3.0 | 2.0 | 3.0 | 2.0 |
| 2014 | - | - | 143 | 40 | 2.5 | 2.5 | 2.5 | 2.5 | 3.0 | 2.0 | 3.0 | 2.0 |

UK (Scotland)-East

Annual input data for NEAC PFA run-reconstruction & NCL models for SCOTLAND (EAST) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
|------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| 1971 | 216,873 | 135,530 | 25.0 | 10.0 | 25.0 | 10.0 | 75.4 | 12.6 | 49.9 | 10.0 |
| 1972 | 220,106 | 183,875 | 25.0 | 10.0 | 25.0 | 10.0 | 76.8 | 12.8 | 51.4 | 10.3 |
| 1973 | 259,773 | 204,826 | 25.0 | 10.0 | 25.0 | 10.0 | 74.9 | 12.5 | 49.9 | 10.0 |
| 1974 | 245,424 | 158,959 | 25.0 | 10.0 | 25.0 | 10.0 | 82.0 | 13.7 | 56.3 | 11.3 |
| 1975 | 181,940 | 180,828 | 25.0 | 10.0 | 25.0 | 10.0 | 80.5 | 13.4 | 55.1 | 11.0 |
| 1976 | 150,069 | 92,179 | 25.0 | 10.0 | 25.0 | 10.0 | 76.5 | 12.8 | 50.7 | 10.1 |
| 1977 | 154,306 | 118,645 | 25.0 | 10.0 | 25.0 | 10.0 | 81.4 | 13.6 | 55.8 | 11.2 |
| 1978 | 158,859 | 139,763 | 25.0 | 10.0 | 25.0 | 10.0 | 75.6 | 12.6 | 51.0 | 10.2 |
| 1979 | 160,796 | 116,559 | 25.0 | 10.0 | 25.0 | 10.0 | 78.4 | 13.1 | 53.9 | 10.8 |
| 1980 | 101,665 | 155,646 | 17.5 | 7.5 | 17.5 | 7.5 | 76.8 | 12.8 | 52.0 | 10.4 |
| 1981 | 129,690 | 156,683 | 17.5 | 7.5 | 17.5 | 7.5 | 75.9 | 12.7 | 51.2 | 10.2 |
| 1982 | 175,374 | 113,198 | 17.5 | 7.5 | 17.5 | 7.5 | 71.1 | 11.8 | 45.3 | 9.1 |
| 1983 | 170,843 | 126,104 | 17.5 | 7.5 | 17.5 | 7.5 | 77.0 | 12.8 | 49.4 | 9.9 |
| 1984 | 175,675 | 90,829 | 17.5 | 7.5 | 17.5 | 7.5 | 70.1 | 11.7 | 43.9 | 8.8 |
| 1985 | 133,119 | 95,044 | 17.5 | 7.5 | 17.5 | 7.5 | 61.9 | 10.3 | 38.9 | 7.8 |
| 1986 | 180,292 | 128,654 | 17.5 | 7.5 | 17.5 | 7.5 | 59.5 | 9.9 | 37.6 | 7.5 |
| 1987 | 139,252 | 88,519 | 17.5 | 7.5 | 17.5 | 7.5 | 64.5 | 10.8 | 40.5 | 8.1 |
| 1988 | 118,614 | 91,151 | 17.5 | 7.5 | 17.5 | 7.5 | 40.3 | 6.7 | 29.2 | 5.8 |
| 1989 | 143,049 | 85,385 | 10.0 | 5.0 | 10.0 | 5.0 | 37.5 | 6.3 | 28.0 | 5.6 |
| 1990 | 63,318 | 73,971 | 10.0 | 5.0 | 10.0 | 5.0 | 39.8 | 6.6 | 28.7 | 5.7 |
| 1991 | 53,860 | 53,693 | 10.0 | 5.0 | 10.0 | 5.0 | 36.8 | 6.1 | 27.4 | 5.5 |
| 1992 | 79,883 | 67,968 | 10.0 | 5.0 | 10.0 | 5.0 | 32.1 | 5.4 | 25.9 | 5.2 |
| 1993 | 73,396 | 60,496 | 10.0 | 5.0 | 10.0 | 5.0 | 35.3 | 5.9 | 26.9 | 5.4 |
| 1994 | 80,429 | 72,758 | 10.0 | 5.0 | 10.0 | 5.0 | 33.1 | 5.5 | 26.1 | 5.2 |
| 1995 | 72,973 | 69,051 | 10.0 | 5.0 | 10.0 | 5.0 | 30.9 | 5.2 | 25.4 | 5.1 |
| 1996 | 56,627 | 50,365 | 10.0 | 5.0 | 10.0 | 5.0 | 28.8 | 4.8 | 24.5 | 4.9 |
| 1997 | 37,448 | 34,850 | 10.0 | 5.0 | 10.0 | 5.0 | 30.6 | 5.1 | 25.1 | 5.0 |
| 1998 | 44,952 | 32,231 | 10.0 | 5.0 | 10.0 | 5.0 | 24.2 | 4.0 | 22.9 | 4.6 |
| 1999 | 20,907 | 27,011 | 10.0 | 5.0 | 10.0 | 5.0 | 24.8 | 4.1 | 23.3 | 4.7 |
| 2000 | 36,871 | 31,280 | 10.0 | 5.0 | 10.0 | 5.0 | 21.8 | 3.6 | 22.3 | 4.5 |
| 2001 | 36,646 | 30,470 | 10.0 | 5.0 | 10.0 | 5.0 | 20.4 | 3.4 | 21.6 | 4.5 |
| 2002 | 26,616 | 21,740 | 10.0 | 5.0 | 10.0 | 5.0 | 19.3 | 3.2 | 21.2 | 4.2 |
| 2003 | 25,871 | 24,270 | 10.0 | 5.0 | 10.0 | 5.0 | 17.3 | 2.8 | 19.3 | 4.3 |
| 2004 | 31,667 | 30,773 | 10.0 | 5.0 | 10.0 | 5.0 | 17.3 | 2.8 | 19.3 | 4.3 |
| 2005 | 31,597 | 23,676 | 10.0 | 5.0 | 10.0 | 5.0 | 17.3 | 2.8 | 19.3 | 4.3 |
| 2006 | 30,739 | 22,954 | 10.0 | 5.0 | 10.0 | 5.0 | 15.3 | 2.8 | 16.5 | 3.5 |
| 2007 | 26,015 | 19,444 | 10.0 | 5.0 | 10.0 | 5.0 | 13.8 | 2.8 | 15.0 | 3.5 |
| 2008 | 18,586 | 20,757 | 10.0 | 5.0 | 10.0 | 5.0 | 10.8 | 2.8 | 14.0 | 3.5 |
| 2009 | 14,863 | 15,042 | 10.0 | 5.0 | 10.0 | 5.0 | 9.8 | 2.8 | 13.0 | 3.5 |
| 2010 | 28,252 | 22,908 | 10.0 | 5.0 | 10.0 | 5.0 | 9.8 | 2.8 | 13.0 | 3.5 |
| 2011 | 12,485 | 24,213 | 10.0 | 5.0 | 10.0 | 5.0 | 9.3 | 2.8 | 12.5 | 3.5 |
| 2012 | 16,117 | 16,165 | 10.0 | 5.0 | 10.0 | 5.0 | 8.3 | 2.8 | 11.5 | 3.5 |
| 2013 | 18,400 | 14,901 | 10.0 | 5.0 | 10.0 | 5.0 | 7.3 | 2.8 | 11.0 | 3.5 |
| 2014 | 10,917 | 11,780 | 10.0 | 5.0 | 10.0 | 5.0 | 5.8 | 2.3 | 9.5 | 3.0 |

UK (Scotland)-West

| Annual input data for NEAC PFA run-reconstruction & NCL models for SCOTLAND (WEST) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation) | | | | | | | | | | |
|-------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|---------------------------|---------------------------|--------------------------------------------|-------------------------------------------------|--------------------------------------------|-------------------------------------------------|----------------------------------------------|---------------------------------------------------|----------------------------------------------|---------------------------------------------------|
| Year | Declared catch 1SW salmon | Declared catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | Estimated % unreported catch of MSW salmon | Uncertainty in % unreported catch of MSW salmon | Estimated exploitation rate (%) - 1SW salmon | Uncertainty in exploitation rate (%) - 1SW salmon | Estimated exploitation rate (%) - MSW salmon | Uncertainty in exploitation rate (%) - MSW salmon |
| 1971 | 45,287 | 26,071 | 35.0 | 10.0 | 35.0 | 10.0 | 37.7 | 6.3 | 24.9 | 5.0 |
| 1972 | 31,358 | 34,148 | 35.0 | 10.0 | 35.0 | 10.0 | 38.4 | 6.4 | 25.7 | 5.1 |
| 1973 | 33,317 | 33,094 | 35.0 | 10.0 | 35.0 | 10.0 | 37.5 | 6.2 | 24.9 | 5.0 |
| 1974 | 43,992 | 29,369 | 35.0 | 10.0 | 35.0 | 10.0 | 41.0 | 6.8 | 28.2 | 5.6 |
| 1975 | 40,424 | 27,145 | 35.0 | 10.0 | 35.0 | 10.0 | 40.2 | 6.7 | 27.5 | 5.5 |
| 1976 | 38,409 | 22,367 | 35.0 | 10.0 | 35.0 | 10.0 | 38.3 | 6.4 | 25.3 | 5.1 |
| 1977 | 39,952 | 20,335 | 35.0 | 10.0 | 35.0 | 10.0 | 40.7 | 6.8 | 27.9 | 5.6 |
| 1978 | 45,611 | 23,191 | 35.0 | 10.0 | 35.0 | 10.0 | 37.8 | 6.3 | 25.5 | 5.1 |
| 1979 | 26,440 | 15,950 | 35.0 | 10.0 | 35.0 | 10.0 | 39.2 | 6.5 | 26.9 | 5.4 |
| 1980 | 19,776 | 16,942 | 27.5 | 7.5 | 27.5 | 7.5 | 38.4 | 6.4 | 26.0 | 5.2 |
| 1981 | 21,048 | 18,038 | 27.5 | 7.5 | 27.5 | 7.5 | 38.0 | 6.3 | 25.6 | 5.1 |
| 1982 | 32,687 | 15,044 | 27.5 | 7.5 | 27.5 | 7.5 | 35.5 | 5.9 | 22.6 | 4.5 |
| 1983 | 38,774 | 19,857 | 27.5 | 7.5 | 27.5 | 7.5 | 38.5 | 6.4 | 24.7 | 4.9 |
| 1984 | 37,404 | 16,384 | 27.5 | 7.5 | 27.5 | 7.5 | 35.1 | 5.8 | 21.9 | 4.4 |
| 1985 | 24,861 | 19,571 | 27.5 | 7.5 | 27.5 | 7.5 | 30.9 | 5.2 | 19.5 | 3.9 |
| 1986 | 22,546 | 19,543 | 27.5 | 7.5 | 27.5 | 7.5 | 29.7 | 5.0 | 18.8 | 3.8 |
| 1987 | 25,533 | 15,475 | 27.5 | 7.5 | 27.5 | 7.5 | 32.3 | 5.4 | 20.3 | 4.1 |
| 1988 | 30,484 | 21,011 | 27.5 | 7.5 | 27.5 | 7.5 | 20.1 | 3.4 | 14.6 | 2.9 |
| 1989 | 31,892 | 18,501 | 20.0 | 5.0 | 20.0 | 5.0 | 18.8 | 3.1 | 14.0 | 2.8 |
| 1990 | 17,776 | 13,953 | 20.0 | 5.0 | 20.0 | 5.0 | 19.9 | 3.3 | 14.4 | 2.9 |
| 1991 | 19,748 | 11,500 | 20.0 | 5.0 | 20.0 | 5.0 | 18.4 | 3.1 | 13.7 | 2.7 |
| 1992 | 21,793 | 14,873 | 20.0 | 5.0 | 20.0 | 5.0 | 16.1 | 2.7 | 12.9 | 2.6 |
| 1993 | 21,121 | 11,230 | 20.0 | 5.0 | 20.0 | 5.0 | 17.7 | 2.9 | 13.5 | 2.7 |
| 1994 | 18,234 | 12,304 | 20.0 | 5.0 | 20.0 | 5.0 | 16.5 | 2.8 | 13.0 | 2.6 |
| 1995 | 16,831 | 9,137 | 20.0 | 5.0 | 20.0 | 5.0 | 15.5 | 2.6 | 12.7 | 2.5 |
| 1996 | 9,537 | 7,463 | 20.0 | 5.0 | 20.0 | 5.0 | 14.4 | 2.4 | 12.2 | 2.4 |
| 1997 | 9,059 | 5,504 | 20.0 | 5.0 | 20.0 | 5.0 | 15.3 | 2.5 | 12.6 | 2.5 |
| 1998 | 8,369 | 6,150 | 20.0 | 5.0 | 20.0 | 5.0 | 12.1 | 2.0 | 11.5 | 2.3 |
| 1999 | 4,147 | 3,587 | 20.0 | 5.0 | 20.0 | 5.0 | 12.4 | 2.1 | 11.7 | 2.3 |
| 2000 | 6,974 | 5,301 | 20.0 | 5.0 | 20.0 | 5.0 | 10.9 | 1.8 | 11.1 | 2.2 |
| 2001 | 5,603 | 4,191 | 20.0 | 5.0 | 20.0 | 5.0 | 10.2 | 1.7 | 10.8 | 2.3 |
| 2002 | 4,691 | 4,548 | 20.0 | 5.0 | 20.0 | 5.0 | 9.6 | 1.6 | 10.6 | 2.1 |
| 2003 | 3,536 | 3,061 | 20.0 | 5.0 | 20.0 | 5.0 | 4.8 | 0.8 | 5.3 | 1.3 |
| 2004 | 5,836 | 6,024 | 20.0 | 5.0 | 20.0 | 5.0 | 7.0 | 1.0 | 7.5 | 1.5 |
| 2005 | 7,428 | 4,913 | 20.0 | 5.0 | 20.0 | 5.0 | 7.0 | 1.0 | 7.5 | 1.5 |
| 2006 | 5,767 | 4,403 | 20.0 | 5.0 | 20.0 | 5.0 | 7.0 | 1.0 | 7.5 | 1.5 |
| 2007 | 6,178 | 4,470 | 20.0 | 5.0 | 20.0 | 5.0 | 7.0 | 1.0 | 7.5 | 1.5 |
| 2008 | 4,740 | 4,853 | 20.0 | 5.0 | 20.0 | 5.0 | 7.0 | 1.0 | 7.5 | 1.5 |
| 2009 | 3,250 | 4,095 | 20.0 | 5.0 | 20.0 | 5.0 | 6.0 | 1.0 | 6.5 | 1.5 |
| 2010 | 5,107 | 4,052 | 20.0 | 5.0 | 20.0 | 5.0 | 6.0 | 1.0 | 6.5 | 1.5 |
| 2011 | 3,206 | 4,246 | 20.0 | 5.0 | 20.0 | 5.0 | 5.5 | 1.0 | 6.0 | 1.5 |
| 2012 | 3,239 | 3,392 | 20.0 | 5.0 | 20.0 | 5.0 | 4.5 | 1.0 | 5.0 | 1.5 |
| 2013 | 2,342 | 2,286 | 20.0 | 5.0 | 20.0 | 5.0 | 4.0 | 1.0 | 4.8 | 1.5 |
| 2014 | 1,651 | 1,466 | 20.0 | 5.0 | 20.0 | 5.0 | 3.5 | 1.0 | 4.5 | 1.5 |

Faroes

Annual input data for NEAC PFA run-reconstruction & NCL models for Faroes. (Uncertainty values define uniform distribution around estimate used in Monte Carlo simulation)

| Year | Catch 1SW salmon | Catch MSW salmon | Estimated % unreported catch of 1SW salmon | Uncertainty in % unreported catch of 1SW salmon | % wild | Natural mortality after 1st sea winter (M) |
|------|------------------|------------------|--------------------------------------------|-------------------------------------------------|--------|--------------------------------------------|
| 1971 | 2,620 | 105,796 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1972 | 2,754 | 111,187 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1973 | 3,121 | 126,012 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1974 | 2,186 | 88,276 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1975 | 2,798 | 112,984 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1976 | 1,830 | 73,900 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1977 | 1,291 | 52,112 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1978 | 974 | 39,309 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1979 | 1,736 | 70,082 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1980 | 4,523 | 182,616 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1981 | 7,443 | 300,542 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1982 | 6,859 | 276,957 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1983 | 15,861 | 215,349 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1984 | 5,534 | 138,227 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1985 | 378 | 158,103 | 10.0 | 5.0 | 0.9 | 0.03 |
| 1986 | 1,979 | 180,934 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1987 | 90 | 166,244 | 10.0 | 5.0 | 1.0 | 0.03 |
| 1988 | 8,637 | 87,629 | 10.0 | 5.0 | 0.9 | 0.03 |
| 1989 | 1,788 | 121,965 | 10.0 | 5.0 | 0.8 | 0.03 |
| 1990 | 1,989 | 140,054 | 10.0 | 5.0 | 0.5 | 0.03 |
| 1991 | 943 | 84,935 | 10.0 | 5.0 | 0.5 | 0.03 |
| 1992 | 68 | 35,700 | 10.0 | 5.0 | 0.6 | 0.03 |
| 1993 | 6 | 30,023 | 10.0 | 5.0 | 0.7 | 0.03 |
| 1994 | 15 | 31,672 | 10.0 | 5.0 | 0.7 | 0.03 |
| 1995 | 18 | 34,662 | 10.0 | 5.0 | 0.8 | 0.03 |
| 1996 | 101 | 28,381 | 10.0 | 5.0 | 0.8 | 0.03 |
| 1997 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 1998 | 339 | 1,424 | 15.0 | 5.0 | 0.8 | 0.03 |
| 1999 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2000 | 225 | 1,765 | 15.0 | 5.0 | 0.8 | 0.03 |
| 2001 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2002 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2003 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2004 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2005 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2006 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2007 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2008 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2009 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2010 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2011 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2012 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2013 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |
| 2014 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.03 |

West Greenland

Annual input data for NEAC PFA run-reconstruction & NCL models for WEST GREENLAND. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| Year | Declared catch (t) | Estimated unreported catch | Wean weight | Estimated min' proportion of NAC fish (from scale analysis) | Estimated max' proportion of NAC fish (from scale analysis) | Proportion 1SW in NAC fish | Proportion 1SW in NEAC fish | No. Fish identified as NAC (from genetic analysis) | No. Fish identified as NEAC (from genetic analysis) |
|------|--------------------|----------------------------|-------------|-------------------------------------------------------------|-------------------------------------------------------------|----------------------------|-----------------------------|----------------------------------------------------|-----------------------------------------------------|
| 1971 | 2689 | 0 | 3.14 | 0.28 | 0.4 | 0.945 | 0.964 | - | - |
| 1972 | 2113 | 0 | 3.44 | 0.34 | 0.37 | 0.945 | 0.964 | - | - |
| 1973 | 2341 | 0 | 4.18 | 0.39 | 0.59 | 0.945 | 0.964 | - | - |
| 1974 | 1917 | 0 | 3.58 | 0.39 | 0.46 | 0.945 | 0.964 | - | - |
| 1975 | 2030 | 0 | 3.12 | 0.4 | 0.48 | 0.945 | 0.964 | - | - |
| 1976 | 1175 | 0 | 3.04 | 0.38 | 0.48 | 0.945 | 0.964 | - | - |
| 1977 | 1420 | 0 | 3.21 | 0.38 | 0.57 | 0.945 | 0.964 | - | - |
| 1978 | 984 | 0 | 3.35 | 0.47 | 0.57 | 0.945 | 0.964 | - | - |
| 1979 | 1395 | 0 | 3.34 | 0.48 | 0.52 | 0.945 | 0.964 | - | - |
| 1980 | 1194 | 0 | 3.22 | 0.45 | 0.51 | 0.945 | 0.964 | - | - |
| 1981 | 1264 | 0 | 3.17 | 0.58 | 0.61 | 0.945 | 0.964 | - | - |
| 1982 | 1077 | 0 | 3.11 | 0.6 | 0.64 | 0.945 | 0.964 | - | - |
| 1983 | 310 | 0 | 3.1 | 0.38 | 0.41 | 0.945 | 0.964 | - | - |
| 1984 | 297 | 0 | 3.11 | 0.47 | 0.53 | 0.945 | 0.964 | - | - |
| 1985 | 864 | 0 | 2.87 | 0.46 | 0.53 | 0.925 | 0.950 | - | - |
| 1986 | 960 | 0 | 3.03 | 0.48 | 0.66 | 0.951 | 0.975 | - | - |
| 1987 | 966 | 0 | 3.16 | 0.54 | 0.63 | 0.963 | 0.980 | - | - |
| 1988 | 893 | 0 | 3.18 | 0.38 | 0.49 | 0.967 | 0.981 | - | - |
| 1989 | 337 | 0 | 2.87 | 0.52 | 0.6 | 0.923 | 0.955 | - | - |
| 1990 | 274 | 0 | 2.69 | 0.7 | 0.79 | 0.957 | 0.963 | - | - |
| 1991 | 472 | 0 | 2.65 | 0.61 | 0.69 | 0.956 | 0.934 | - | - |
| 1992 | 237 | 0 | 2.81 | 0.5 | 0.57 | 0.919 | 0.975 | - | - |
| 1993 | 0 | 12 | 2.73 | 0.5 | 0.76 | 0.946 | 0.961 | - | - |
| 1994 | 0 | 12 | 2.73 | 0.5 | 0.76 | 0.946 | 0.961 | - | - |
| 1995 | 83 | 20 | 2.56 | 0.65 | 0.72 | 0.968 | 0.973 | - | - |
| 1996 | 92 | 20 | 2.88 | 0.71 | 0.76 | 0.941 | 0.961 | - | - |
| 1997 | 58 | 5 | 2.71 | 0.75 | 0.84 | 0.982 | 0.993 | - | - |
| 1998 | 11 | 11 | 2.78 | 0.73 | 0.84 | 0.968 | 0.994 | - | - |
| 1999 | 19 | 12.5 | 3.08 | 0.84 | 0.97 | 0.968 | 1.000 | - | - |
| 2000 | 21 | 10 | 2.57 | 0 | 0 | 0.974 | 1.000 | 344 | 146 |
| 2001 | 43 | 10 | 3 | 0.67 | 0.71 | 0.982 | 0.978 | 1 | 1 |
| 2002 | 9.8 | 10 | 2.9 | 0 | 0 | 0.973 | 1.000 | 338 | 163 |
| 2003 | 12.3 | 10 | 3.04 | 0 | 0 | 0.967 | 0.989 | 1,212 | 567 |
| 2004 | 17.2 | 10 | 3.18 | 0 | 0 | 0.970 | 0.970 | 1,192 | 447 |
| 2005 | 17.3 | 10 | 3.31 | 0 | 0 | 0.924 | 0.967 | 585 | 182 |
| 2006 | 23 | 10 | 3.24 | 0 | 0 | 0.930 | 0.988 | 857 | 326 |
| 2007 | 24.8 | 10 | 2.98 | 0 | 0 | 0.965 | 0.956 | 917 | 206 |
| 2008 | 28.6 | 10 | 3.08 | 0 | 0 | 0.974 | 0.988 | 1,593 | 260 |
| 2009 | 28 | 10 | 3.5 | 0 | 0 | 0.934 | 0.894 | 1,483 | 138 |
| 2010 | 43.1 | 10 | 3.42 | 0 | 0 | 0.982 | 0.975 | 991 | 249 |
| 2011 | 27.4 | 10 | 3.4 | 0 | 0 | 0.939 | 0.831 | 888 | 72 |
| 2012 | 34.6 | 10 | 3.44 | 0 | 1 | 0.932 | 0.980 | 1,121 | 252 |
| 2013 | 47.7 | 10 | 3.35 | 0 | 1 | 0.949 | 0.966 | 938 | 211 |
| 2014 | 58.2 | 10 | 3.32 | 0 | 1 | 0.913 | 0.961 | 660 | 260 |

| Stock composition | |
|-------------------|-------|
| Country | MSW |
| France | 0.027 |
| Finland | 0.001 |
| Iceland | 0.001 |
| Ireland | 0.147 |
| Norway | 0.027 |
| Russia | 0.000 |
| Sweden | 0.003 |
| UK (E&W) | 0.149 |
| UK (NI) | 0.000 |
| UK (Sc) | 0.645 |
| Other | |
| Total | 1 |

Appendix 4: Input data for Atlantic salmon used to do the run-reconstruction and estimates of returns and spawners by size group and age group for North America

Appendix 4.i. Input data for the fishery at West Greenland used in the run reconstruction model.

| Year of the fishery | Harvest of salmon at West Greenland in tons | Unreported harvest of salmon at West Greenland in tons | Mean weight of salmon (all ages and origin) at West Greenland | Sample size of salmon assigned to NAC based on genetic identification (since 2002) | Sample size of salmon assigned to NEAC based on genetic identification (since 2002) | Lower CI of prop. of salmon assigned to NAC based on scale analyses and discriminant analyses | Upper CI of prop. of salmon assigned to NAC based on scale analyses and discriminant analyses | Prop. of salmon of NAC origin which are 1SW non-maturing age group | Prop. of salmon of NEAC origin which are 1SW non-maturing age group |
|---------------------|---------------------------------------------|--------------------------------------------------------|---------------------------------------------------------------|------------------------------------------------------------------------------------|-------------------------------------------------------------------------------------|-----------------------------------------------------------------------------------------------|-----------------------------------------------------------------------------------------------|--------------------------------------------------------------------|---------------------------------------------------------------------|
| Year | WGHav | WGUHav | WGMeanWt | WGSampleNAC | WGSampleNEAC | WGPropNACMin | WGPropNACMax | WGProp1SWNAC | WGProp1SWNEAC |
| 1970 | 0 | 0 | 3 | 0 | 0 | 0.2 | 0.5 | 0.9 | 1 |
| 1971 | 2689 | 0 | 3.14 | 0 | 0 | 0.28 | 0.4 | 0.945125 | 0.964125 |
| 1972 | 2113 | 0 | 3.44 | 0 | 0 | 0.34 | 0.37 | 0.945125 | 0.964125 |
| 1973 | 2341 | 0 | 4.18 | 0 | 0 | 0.39 | 0.59 | 0.945125 | 0.964125 |
| 1974 | 1917 | 0 | 3.58 | 0 | 0 | 0.39 | 0.46 | 0.945125 | 0.964125 |
| 1975 | 2030 | 0 | 3.12 | 0 | 0 | 0.4 | 0.48 | 0.945125 | 0.964125 |
| 1976 | 1175 | 0 | 3.04 | 0 | 0 | 0.38 | 0.48 | 0.945125 | 0.964125 |
| 1977 | 1420 | 0 | 3.2125 | 0 | 0 | 0.38 | 0.57 | 0.945125 | 0.964125 |
| 1978 | 984 | 0 | 3.35 | 0 | 0 | 0.47 | 0.57 | 0.945125 | 0.964125 |
| 1979 | 1395 | 0 | 3.34 | 0 | 0 | 0.48 | 0.52 | 0.945125 | 0.964125 |
| 1980 | 1194 | 0 | 3.22 | 0 | 0 | 0.45 | 0.51 | 0.945125 | 0.964125 |
| 1981 | 1264 | 0 | 3.17 | 0 | 0 | 0.58 | 0.61 | 0.945125 | 0.964125 |
| 1982 | 1077 | 0 | 3.11 | 0 | 0 | 0.6 | 0.64 | 0.945125 | 0.964125 |
| 1983 | 310 | 0 | 3.1 | 0 | 0 | 0.38 | 0.41 | 0.945125 | 0.964125 |
| 1984 | 297 | 0 | 3.11 | 0 | 0 | 0.47 | 0.53 | 0.945125 | 0.964125 |
| 1985 | 864 | 0 | 2.87 | 0 | 0 | 0.46 | 0.53 | 0.925 | 0.95 |
| 1986 | 960 | 0 | 3.03 | 0 | 0 | 0.48 | 0.66 | 0.951 | 0.975 |
| 1987 | 966 | 0 | 3.16 | 0 | 0 | 0.54 | 0.63 | 0.963 | 0.98 |
| 1988 | 893 | 0 | 3.18 | 0 | 0 | 0.38 | 0.49 | 0.967 | 0.981 |
| 1989 | 337 | 0 | 2.87 | 0 | 0 | 0.52 | 0.6 | 0.923 | 0.955 |
| 1990 | 274 | 0 | 2.69 | 0 | 0 | 0.7 | 0.79 | 0.957 | 0.963 |
| 1991 | 472 | 0 | 2.65 | 0 | 0 | 0.61 | 0.69 | 0.956 | 0.934 |
| 1992 | 237 | 0 | 2.81 | 0 | 0 | 0.5 | 0.57 | 0.919 | 0.975 |
| 1993 | 0 | 12 | 2.725 | 0 | 0 | 0.5 | 0.76 | 0.946 | 0.96075 |
| 1994 | 0 | 12 | 2.725 | 0 | 0 | 0.5 | 0.76 | 0.946 | 0.96075 |
| 1995 | 83 | 20 | 2.56 | 0 | 0 | 0.65 | 0.72 | 0.968 | 0.973 |
| 1996 | 92 | 20 | 2.88 | 0 | 0 | 0.71 | 0.76 | 0.941 | 0.961 |
| 1997 | 58 | 5 | 2.71 | 0 | 0 | 0.75 | 0.84 | 0.982 | 0.993 |
| 1998 | 11 | 11 | 2.78 | 0 | 0 | 0.73 | 0.84 | 0.968 | 0.994 |
| 1999 | 19 | 12.5 | 3.08 | 0 | 0 | 0.84 | 0.97 | 0.968 | 1 |
| 2000 | 21 | 10 | 2.57 | 344 | 146 | 0 | 0 | 0.974 | 1 |
| 2001 | 43 | 10 | 3 | 1 | 1 | 0.67 | 0.71 | 0.982 | 0.978 |
| 2002 | 9.8 | 10 | 2.9 | 338 | 163 | 0 | 0 | 0.973 | 1 |
| 2003 | 12.3 | 10 | 3.04 | 1212 | 567 | 0 | 0 | 0.967 | 0.989 |
| 2004 | 17.2 | 10 | 3.18 | 1192 | 447 | 0 | 0 | 0.97 | 0.97 |
| 2005 | 17.3 | 10 | 3.31 | 585 | 182 | 0 | 0 | 0.924 | 0.967 |
| 2006 | 23 | 10 | 3.24 | 857 | 326 | 0 | 0 | 0.93 | 0.988 |
| 2007 | 24.8 | 10 | 2.98 | 917 | 206 | 0 | 0 | 0.965 | 0.956 |
| 2008 | 28.6 | 10 | 3.08 | 1593 | 260 | 0 | 0 | 0.974 | 0.988 |
| 2009 | 28 | 10 | 3.5 | 1483 | 138 | 0 | 0 | 0.934 | 0.894 |
| 2010 | 43.1 | 10 | 3.42 | 991 | 249 | 0 | 0 | 0.982 | 0.975 |
| 2011 | 27.4 | 10 | 3.4 | 888 | 72 | 0 | 0 | 0.939 | 0.831 |
| 2012 | 34.5 | 10 | 3.44 | 1121 | 252 | 0 | 0 | 0.932 | 0.98 |
| 2013 | 47.7 | 10 | 3.35 | 938 | 211 | 0 | 0 | 0.949 | 0.966 |
| 2014 | 58.282 | 10 | 3.32 | 660 | 260 | 0 | 0 | 0.913 | 0.961 |

Appendix 4.ii. Input data for sea fisheries on large salmon and small salmon from Newfoundland and Labrador used in the run reconstruction model. Labrador represents harvests from Labrador in aboriginal fisheries for food, social and ceremonial purposes and the resident food fishery beginning in 1998.

| Year of the fishery | Catches of large salmon | | | Catches of small salmon | | |
|---------------------|-------------------------|--------------|----------------------|-------------------------|--------------|--------------|
| | SFA 1 to 7 | SFA 8 to 14A | Subsistence Labrador | SFA 1 to 7 | SFA 8 to 14A | FSC Labrador |
| Winbugs labels | Nlg_LBandNF1to7 | Nlg_NF8to14a | Nlg_LBFSC | Nsm_LBandNF1to7 | Nsm_NF8to14a | Nsm_LBFSC |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 199176 | 0 | 0 | 158896 | 70936 | 0 |
| 1972 | 144496 | 42861 | 0 | 143232 | 111141 | 0 |
| 1973 | 227779 | 43627 | 0 | 188725 | 176907 | 0 |
| 1974 | 196726 | 85714 | 0 | 192195 | 153278 | 0 |
| 1975 | 215025 | 72814 | 0 | 302348 | 91935 | 0 |
| 1976 | 210858 | 95714 | 0 | 221766 | 118779 | 0 |
| 1977 | 231393 | 63449 | 0 | 220093 | 57472 | 0 |
| 1978 | 155546 | 37653 | 0 | 102403 | 38180 | 0 |
| 1979 | 82174 | 29122 | 0 | 186558 | 62622 | 0 |
| 1980 | 211896 | 54307 | 0 | 290127 | 94291 | 0 |
| 1981 | 211006 | 38663 | 0 | 288902 | 60668 | 0 |
| 1982 | 129319 | 35055 | 0 | 222894 | 77017 | 0 |
| 1983 | 108430 | 28215 | 0 | 166033 | 55683 | 0 |
| 1984 | 87742 | 15135 | 0 | 123774 | 52813 | 0 |
| 1985 | 70970 | 24383 | 0 | 178719 | 79275 | 0 |
| 1986 | 107561 | 22036 | 0 | 222671 | 91912 | 0 |
| 1987 | 146242 | 19241 | 0 | 281762 | 82401 | 0 |
| 1988 | 86047 | 14763 | 0 | 198484 | 74620 | 0 |
| 1989 | 85319 | 15577 | 0 | 172861 | 60884 | 0 |
| 1990 | 59334 | 11639 | 0 | 104788 | 46053 | 0 |
| 1991 | 39257 | 10259 | 0 | 89099 | 42721 | 0 |
| 1992 | 32341 | 0 | 0 | 24249 | 0 | 0 |
| 1993 | 17096 | 0 | 0 | 17074 | 0 | 0 |
| 1994 | 15377 | 0 | 0 | 8640 | 0 | 0 |
| 1995 | 11176 | 0 | 0 | 7980 | 0 | 0 |
| 1996 | 7272 | 0 | 0 | 7849 | 0 | 0 |
| 1997 | 6943 | 0 | 0 | 9753 | 0 | 0 |
| 1998 | 0 | 0 | 2269 | 0 | 0 | 2988 |
| 1999 | 0 | 0 | 1084 | 0 | 0 | 2739 |
| 2000 | 0 | 0 | 1352 | 0 | 0 | 5323 |
| 2001 | 0 | 0 | 1721 | 0 | 0 | 4789 |
| 2002 | 0 | 0 | 1389 | 0 | 0 | 5806 |
| 2003 | 0 | 0 | 2175 | 0 | 0 | 6477 |
| 2004 | 0 | 0 | 3696 | 0 | 0 | 8385 |
| 2005 | 0 | 0 | 2817 | 0 | 0 | 10436 |
| 2006 | 0 | 0 | 3090 | 0 | 0 | 10377 |
| 2007 | 0 | 0 | 2652 | 0 | 0 | 9208 |
| 2008 | 0 | 0 | 3909 | 0 | 0 | 9834 |
| 2009 | 0 | 0 | 3344 | 0 | 0 | 7988 |
| 2010 | 0 | 0 | 3725 | 0 | 0 | 9867 |
| 2011 | 0 | 0 | 4451 | 0 | 0 | 11138 |
| 2012 | 0 | 0 | 4228 | 0 | 0 | 9977 |
| 2013 | 0 | 0 | 6375 | 0 | 0 | 7164 |
| 2014 | 0 | 0 | 4003 | 0 | 0 | 8965 |

Appendix 4.iii. Input data for sea fisheries on large salmon and small salmon from St Pierre & Miquelon used in the run-reconstruction model.

| Year of the fishery | Reported harvest (kg) | Number of salmon | Number of large salmon | Number of small salmon |
|---------------------|-----------------------|------------------|------------------------|------------------------|
| Winbugs labels | | | SPMNLarge[] | SPMNSmall[] |
| 1970 | 0 | 0 | 0 | 0 |
| 1971 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 |
| 1976 | 3000 | 1331 | 333 | 998 |
| 1977 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 |
| 1983 | 3000 | 1331 | 333 | 998 |
| 1984 | 3000 | 1331 | 333 | 998 |
| 1985 | 3000 | 1331 | 333 | 998 |
| 1986 | 2500 | 1109 | 277 | 832 |
| 1987 | 2000 | 887 | 222 | 665 |
| 1988 | 2000 | 887 | 222 | 665 |
| 1989 | 2000 | 887 | 222 | 665 |
| 1990 | 1900 | 843 | 211 | 632 |
| 1991 | 1200 | 532 | 133 | 399 |
| 1992 | 2300 | 1020 | 255 | 765 |
| 1993 | 2900 | 1287 | 322 | 965 |
| 1994 | 3400 | 1508 | 377 | 1131 |
| 1995 | 800 | 355 | 89 | 266 |
| 1996 | 1600 | 710 | 177 | 532 |
| 1997 | 1500 | 665 | 166 | 499 |
| 1998 | 2300 | 1020 | 255 | 765 |
| 1999 | 2322 | 1030 | 258 | 773 |
| 2000 | 2267 | 1006 | 251 | 754 |
| 2001 | 2155 | 956 | 239 | 717 |
| 2002 | 1952 | 866 | 217 | 650 |
| 2003 | 2892 | 1283 | 321 | 962 |
| 2004 | 2784 | 1235 | 309 | 926 |
| 2005 | 3287 | 1458 | 365 | 1094 |
| 2006 | 3555 | 1577 | 394 | 1183 |
| 2007 | 1947 | 864 | 216 | 648 |
| 2008 | 3540 | 1571 | 393 | 1178 |
| 2009 | 3460 | 1535 | 384 | 1151 |
| 2010 | 2780 | 1233 | 308 | 925 |
| 2011 | 3757 | 1667 | 417 | 1250 |
| 2012 | 1450 | 643 | 161 | 482 |
| 2013 | 5300 | 2351 | 588 | 1764 |
| 2014 | 3810 | 1690 | 423 | 1268 |

Appendix 4.iv. Input data for large salmon for Labrador used in the run-reconstruction.

| Year | Large Salmon | | | FSC Fishery | Proportion Labrador origin | | | | | | Exploitation rate | | Proportion 2SW | | Returns to Labrador rivers | | Angling catches | | |
|-------------------|-------------------------------------|-------------------------|--------------------------|-------------|----------------------------|-----------------------|-----------------------|-----------------------|-------------------------|-------------------------|--------------------|--------------------|----------------|----------|----------------------------|-----------|-----------------------|-----------------------|----------|
| | Commercial harvest (number of fish) | | | | | | | | | | | | | | number of fish | | number of fish | | |
| | SFA 1 | SFA 2 | SFA 14B | | ALL | SFA 1 | | SFA 2 | | SFA 14B | | All SFAs | | | | | | | |
| | | | | | Number of fish | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Retained | Released |
| 1970 | 25127 | 64806 | 13673 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 562 | 0 | |
| 1971 | 21599 | 55708 | 11753 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 486 | 0 | |
| 1972 | 30204 | 77902 | 16436 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 424 | 0 | |
| 1973 | 13866 | 93036 | 15863 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 1009 | 0 | |
| 1974 | 28601 | 71168 | 14752 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 803 | 0 | |
| 1975 | 38555 | 77796 | 15189 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 327 | 0 | |
| 1976 | 28158 | 70158 | 18664 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 830 | 0 | |
| 1977 | 30824 | 48934 | 11715 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 1286 | 0 | |
| 1978 | 21291 | 27073 | 3874 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 767 | 0 | |
| 1979 | 28750 | 87067 | 9138 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 609 | 0 | |
| 1980 | 36147 | 68581 | 7606 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 889 | 0 | |
| 1981 | 24192 | 53085 | 5966 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 520 | 0 | |
| 1982 | 19403 | 33320 | 7489 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 621 | 0 | |
| 1983 | 11726 | 25258 | 6218 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 428 | 0 | |
| 1984 | 13252 | 16789 | 3954 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 510 | 0 | |
| 1985 | 19152 | 34071 | 5342 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 294 | 0 | |
| 1986 | 18257 | 49799 | 11114 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 467 | 0 | |
| 1987 | 12621 | 32386 | 4591 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 633 | 0 | |
| 1988 | 16261 | 26836 | 4646 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 710 | 0 | |
| 1989 | 7313 | 17316 | 2858 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 461 | 0 | |
| 1990 | 1369 | 7679 | 4417 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 357 | 0 | |
| 1991 | 9981 | 19608 | 2752 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.580 | 0.830 | 0.70 | 0.90 | 0 | 0 | 93 | 0 | |
| 1992 | 3825 | 9651 | 3620 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.38 | 0.62 | 0.70 | 0.90 | 0 | 0 | 781 | 10 | |
| 1993 | 3464 | 11056 | 857 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.29 | 0.50 | 0.70 | 0.90 | 0 | 0 | 378 | 91 | |
| 1994 | 2150 | 8714 | 312 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.14 | 0.25 | 0.70 | 0.90 | 0 | 0 | 455 | 347 | |
| 1995 | 1375 | 5479 | 418 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.13 | 0.23 | 0.70 | 0.90 | 0 | 0 | 408 | 508 | |
| 1996 | 1393 | 5550 | 263 | | 0.6433 | 0.7247 | 0.8839 | 0.9521 | 0.6 | 0.8 | 0.17 | 0.30 | 0.70 | 0.90 | 0 | 0 | 334 | 489 | |
| 1997 | 6943 | | | | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 0 | 0 | 158 | 566 | |
| 1998 | 0 | 0 | 0 | 2269 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 7374 | 19486 | 231 | 814 | |
| 1999 | 0 | 0 | 0 | 1084 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 8827 | 23328 | 320 | 931 | |
| 2000 | 0 | 0 | 0 | 1352 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 12052 | 31850 | 262 | 1446 | |
| 2001 | 0 | 0 | 0 | 1721 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 12744 | 33677 | 338 | 1468 | |
| 2002 | 0 | 0 | 0 | 1389 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 9076 | 24769 | 207 | 978 | |
| 2003 | 0 | 0 | 0 | 2175 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 6676 | 21689 | 222 | 1326 | |
| 2004 | 0 | 0 | 0 | 3696 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 10964 | 23092 | 259 | 1519 | |
| 2005 | 0 | 0 | 0 | 2817 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 11159 | 30796 | 291 | 1290 | |
| 2006 | 0 | 0 | 0 | 3090 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 12414 | 29783 | 227 | 1133 | |
| 2007 | 0 | 0 | 0 | 2652 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.71 | 11887 | 31913 | 235 | 1222 | |
| 2008 | 0 | 0 | 0 | 3909 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.30 | 0.60 | 0.70 | 14700 | 37677 | 200 | 1461 | |
| 2009 | 0 | 0 | 0 | 3344 | 1 | 1 | 1 | 1 | 1 | 1 | 0.2 | 0.4 | 0.60 | 0.70 | 18643 | 60062 | 216 | 1219 | |
| 2010 | 0 | 0 | 0 | 3725 | 1 | 1 | 1 | 1 | 1 | 1 | 0.2 | 0.4 | 0.60 | 0.70 | 7498 | 20099 | 197 | 1080 | |
| 2011 | 0 | 0 | 0 | 4451 | 1 | 1 | 1 | 1 | 1 | 1 | 0.2 | 0.4 | 0.60 | 0.70 | 30198 | 85085 | 0 | 2233 | |
| 2012 | 0 | 0 | 0 | 4228 | 1 | 1 | 1 | 1 | 1 | 1 | 0.2 | 0.4 | 0.60 | 0.70 | 19062 | 48538 | 0 | 1072 | |
| 2013 | 0 | 0 | 0 | 6479 | 1 | 1 | 1 | 1 | 1 | 1 | 0.2 | 0.4 | 0.60 | 0.70 | 37971 | 94955 | 0 | 2433 | |
| 2014 | 0 | 0 | 0 | 4003 | 1 | 1 | 1 | 1 | 1 | 1 | 0.2 | 0.4 | 0.60 | 0.70 | 45436 | 109693 | 0 | 1518 | |
| Winbugs variables | LB_SFA1_L [g_Comm[]] | LB_SFA2_L [g_Comm[]] | LB_SFA14B [Lg_Comm[]] | NLg_LBFS[] | pLB_SFA1_ [Lg_L[]] | pLB_SFA1_ [Lg_H[]] | pLB_SFA2_ [Lg_L[]] | pLB_SFA2_ [Lg_H[]] | pLB_SFA14 [B_Lg_L[]] | pLB_SFA14 [B_Lg_H[]] | ER_LB_Lg_ [L[]] | ER_LB_Lg_ [H[]] | p2SW_L[] | p2SW_H[] | LB_Lg_L[] | LB_Lg_H[] | LB_Ang_Lg_ [Ret[]] | LB_Ang_Lg_ [Rel[]] | |

Appendix 4.v. Input data for small salmon for Labrador used in the run-reconstruction.

| Year | Small salmon | | | FSC Fishery ALL Number of fish | Proportion Labrador origin | | | | | | | | Exploitation rate | | Returns to Labrador rivers | | Angling catches | |
|-------------------|-------------------------------------|-------------------|--------------------|--------------------------------------|----------------------------|-----------------|-----------------|-----------------|------------------|------------------|--------------|--------------|-------------------|-----------|----------------------------|-----------------|-----------------|----------|
| | Commercial harvest (number of fish) | | | | SFA 2 | | | | SFA 14B | | | | All SFAs | | number of fish | | number of fish | |
| | SFA 1 | SFA 2 | SFA 14B | | SFA 1 | | SFA 2 | | SFA 14B | | All SFAs | | Min | Max | Min | Max | Retained | Released |
| | | | | | Min | Max | Min | Max | Min | Max | Min | Max | | | | | | |
| 1970 | 19109 | 38359 | 11212 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 4013 | 0 |
| 1971 | 14303 | 28711 | 8392 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 3934 | 0 |
| 1972 | 3130 | 6282 | 1836 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 2947 | 0 |
| 1973 | 9848 | 37145 | 9328 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 7492 | 0 |
| 1974 | 34937 | 57560 | 19294 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 2501 | 0 |
| 1975 | 17589 | 47468 | 13152 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 3972 | 0 |
| 1976 | 17796 | 40539 | 11267 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 5726 | 0 |
| 1977 | 17095 | 12535 | 4026 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 4594 | 0 |
| 1978 | 9712 | 28808 | 7194 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 2691 | 0 |
| 1979 | 22501 | 72485 | 8493 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 4118 | 0 |
| 1980 | 21596 | 86426 | 6658 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 3800 | 0 |
| 1981 | 18478 | 53592 | 7379 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 5191 | 0 |
| 1982 | 15964 | 30185 | 3292 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 4104 | 0 |
| 1983 | 11474 | 11695 | 2421 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 4372 | 0 |
| 1984 | 15400 | 24499 | 7460 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 2935 | 0 |
| 1985 | 17779 | 45321 | 8296 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 3101 | 0 |
| 1986 | 13714 | 64351 | 11389 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 3464 | 0 |
| 1987 | 19641 | 56381 | 7087 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 5366 | 0 |
| 1988 | 13233 | 34200 | 9053 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 5523 | 0 |
| 1989 | 8736 | 20699 | 3592 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 4684 | 0 |
| 1990 | 1410 | 20055 | 5303 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.3 | 0.5 | | | | | 3309 | 0 |
| 1991 | 9588 | 13336 | 1325 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.22 | 0.39 | | | | | 2323 | 0 |
| 1992 | 3893 | 12037 | 1144 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.13 | 0.25 | | | | | 2738 | 251 |
| 1993 | 3303 | 4535 | 802 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.10 | 0.19 | | | | | 2508 | 1793 |
| 1994 | 3202 | 4561 | 217 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.07 | 0.13 | | | | | 2549 | 3681 |
| 1995 | 1676 | 5308 | 865 | | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.04 | 0.07 | | | | | 2493 | 3302 |
| 1996 | 1728 | 8025 | 332 | | 0.3557 | 0.4163 | 0.748 | 0.85 | 0.6 | 0.8 | 0.05 | 0.08 | | | | | 2565 | 3776 |
| 1997 | 9753 | | | | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | | | | | 2365 | 2187 |
| 1998 | 0 | 0 | 0 | 2988 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 97408 | 205197 | 2131 | 3758 | | |
| 1999 | 0 | 0 | 0 | 2739 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 94894 | 199901 | 2076 | 4407 | | |
| 2000 | 0 | 0 | 0 | 5323 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 117063 | 246602 | 2561 | 7095 | | |
| 2001 | 0 | 0 | 0 | 4789 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 93660 | 197301 | 2049 | 4640 | | |
| 2002 | 0 | 0 | 0 | 5806 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 62321 | 142951 | 2071 | 5052 | | |
| 2003 | 0 | 0 | 0 | 6477 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 48256 | 122813 | 2112 | 4924 | | |
| 2004 | 0 | 0 | 0 | 8385 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 69808 | 120244 | 1808 | 5968 | | |
| 2005 | 0 | 0 | 0 | 10436 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 160038 | 281401 | 2007 | 7120 | | |
| 2006 | 0 | 0 | 0 | 10377 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 132205 | 294669 | 1656 | 5815 | | |
| 2007 | 0 | 0 | 0 | 9208 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 131895 | 257360 | 1762 | 4641 | | |
| 2008 | 0 | 0 | 0 | 9834 | 1 | 1 | 1 | 1 | 1 | 1 | 0.05 | 0.08 | 142851 | 264694 | 1936 | 5917 | | |
| 2009 | 0 | 0 | 0 | 7988 | 1 | 1 | 1 | 1 | 1 | 1 | 0.07 | 0.14 | 38031 | 140890 | 1355 | 3396 | | |
| 2010 | 0 | 0 | 0 | 9667 | 1 | 1 | 1 | 1 | 1 | 1 | 0.07 | 0.14 | 55949 | 127622 | 1477 | 4704 | | |
| 2011 | 0 | 0 | 0 | 11138 | 1 | 1 | 1 | 1 | 1 | 1 | 0.07 | 0.14 | 137465 | 356791 | 1628 | 5340 | | |
| 2012 | 0 | 0 | 0 | 9977 | 1 | 1 | 1 | 1 | 1 | 1 | 0.07 | 0.14 | 105443 | 241754 | 1376 | 3302 | | |
| 2013 | 0 | 0 | 0 | 7185 | 1 | 1 | 1 | 1 | 1 | 1 | 0.07 | 0.14 | 103596 | 276818 | 1389 | 4167 | | |
| 2014 | 0 | 0 | 0 | 8965 | 1 | 1 | 1 | 1 | 1 | 1 | 0.07 | 0.14 | 217586 | 480361 | 1210 | 3924 | | |
| Winbugs variables | LB_SFA1_Sm_Comm[] | LB_SFA2_Sm_Comm[] | LB_SFA14_Sm_Comm[] | NSm_LBFS_C[] | pLB_SFA1_Sm_L[] | pLB_SFA1_Sm_H[] | pLB_SFA2_Sm_L[] | pLB_SFA2_Sm_H[] | pLB_SFA14_Sm_L[] | pLB_SFA14_Sm_H[] | ER_LB_Sm_L[] | ER_LB_Sm_H[] | LB_Sm_L[] | LB_Sm_H[] | LB_Ang_Sm_Rel[] | LB_Ang_Sm_Rel[] | | |

Appendix 4.vi. Input data for returns of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

| Year | Salmon Fishing Area 3 | | | | | | Salmon Fishing Area 4 | | | | | | Salmon Fishing Area 5 | | | | | | Salmon Fishing Area 6 | | | | | | Salmon Fishing Area 7 | | | | | | Salmon Fishing Area 8 | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
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| | Small salmon | | | Large salmon | | | Small salmon | | | Large salmon | | | Small salmon | | | Large salmon | | | Small salmon | | | Large salmon | | | Small salmon | | | Large salmon | | | Small salmon | | | Large salmon | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| | Returns | | Max | Returns | | Max | Returns | | Max | Returns | | Max | Returns | | Max | Returns | | Max | Returns | | Max | Returns | | Max | Returns | | Max | Returns | | Max | Returns | | Max | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
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| 1970 | 2613 | 5227 | 155 | 737 | 16163 | 32327 | 957 | 4559 | 7420 | 14840 | 439 | 2093 | 280 | 560 | 17 | 79 | 67 | 133 | 4 | 19 | 62 | 123 | 4 | 17 | 1971 | 2473 | 4947 | 146 | 698 | 12610 | 25220 | 746 | 3557 | 5600 | 11200 | 331 | 1579 | 183 | 367 | 11 | 52 | 133 | 267 | 8 | 38 | 83 | 167 | 5 | 24 | 1972 | 1660 | 3320 | 98 | 468 | 11480 | 22960 | 679 | 3238 | 6317 | 12633 | 374 | 1782 | 397 | 793 | 23 | 112 | 203 | 407 | 12 | 57 | 93 | 187 | 6 | 26 | 1973 | 3960 | 7920 | 234 | 1117 | 22367 | 44733 | 1324 | 6308 | 7040 | 14080 | 417 | 1986 | 833 | 1667 | 49 | 235 | 437 | 873 | 26 | 123 | 313 | 627 | 19 | 88 | 1974 | 2797 | 5593 | 322 | 645 | 17910 | 35820 | 2065 | 4131 | 5457 | 10913 | 629 | 1258 | 1010 | 2020 | 116 | 233 | 443 | 887 | 51 | 102 | 170 | 340 | 20 | 39 | 1975 | 3690 | 7380 | 520 | 1041 | 19810 | 39620 | 2794 | 5587 | 6627 | 13253 | 935 | 1869 | 313 | 627 | 44 | 88 | 133 | 267 | 19 | 38 | 290 | 580 | 41 | 82 | 1976 | 3157 | 6313 | 380 | 760 | 22277 | 44553 | 2683 | 5365 | 6327 | 12653 | 762 | 1524 | 823 | 1647 | 99 | 198 | 100 | 200 | 12 | 24 | 267 | 533 | 32 | 64 | 1977 | 5100 | 10200 | 482 | 964 | 27987 | 55973 | 2645 | 5290 | 15387 | 30773 | 1454 | 2908 | 1337 | 2673 | 126 | 253 | 260 | 520 | 25 | 49 | 270 | 540 | 26 | 51 | 1978 | 2527 | 5053 | 150 | 299 | 29247 | 58493 | 1731 | 3461 | 9527 | 19053 | 564 | 1128 | 987 | 1973 | 58 | 117 | 330 | 660 | 20 | 39 | 147 | 293 | 9 | 17 | 1979 | 6800 | 13600 | 390 | 779 | 26753 | 53507 | 1533 | 3067 | 4437 | 8873 | 254 | 509 | 813 | 1627 | 47 | 93 | 417 | 833 | 24 | 48 | 333 | 667 | 19 | 38 | 1980 | 5810 | 11620 | 261 | 522 | 31380 | 62760 | 1410 | 2819 | 9007 | 18013 | 405 | 809 | 1067 | 2133 | 48 | 96 | 340 | 680 | 15 | 31 | 400 | 800 | 18 | 36 | 1981 | 7860 | 15720 | 1045 | 2090 | 45120 | 90240 | 5998 | 11996 | 11627 | 23253 | 1546 | 3091 | 2017 | 4033 | 268 | 536 | 410 | 820 | 55 | 109 | 257 | 513 | 34 | 68 | 1982 | 8780 | 17560 | 212 | 424 | 33243 | 66487 | 802 | 1604 | 8110 | 16220 | 196 | 391 | 960 | 1920 | 23 | 46 | 517 | 1033 | 12 | 25 | 283 | 567 | 7 | 14 | 1983 | 5390 | 10780 | 247 | 495 | 29847 | 59693 | 1370 | 2740 | 7857 | 15713 | 361 | 721 | 987 | 1973 | 45 | 91 | 463 | 927 | 21 | 43 | 137 | 273 | 6 | 13 | 1984 | 3532 | 7526 | 55 | 540 | 34933 | 74436 | 548 | 5337 | 9538 | 20323 | 150 | 1457 | 1101 | 2346 | 17 | 168 | 339 | 722 | 5 | 52 | 279 | 594 | 4 | 43 | 1985 | 4772 | 9879 | 72 | 683 | 44408 | 91931 | 671 | 6352 | 12692 | 26275 | 192 | 1816 | 1563 | 3235 | 24 | 224 | 408 | 845 | 6 | 58 | 375 | 777 | 6 | 54 | 1986 | 2826 | 5898 | 70 | 413 | 34015 | 70993 | 840 | 4977 | 14835 | 30963 | 366 | 2170 | 1629 | 3400 | 40 | 238 | 373 | 779 | 9 | 55 | 505 | 1054 | 12 | 74 | 1987 | 2218 | 4458 | 57 | 318 | 21485 | 43175 | 556 | 3078 | 6556 | 13175 | 170 | 939 | 540 | 1085 | 14 | 77 | 110 | 222 | 3 | 16 | 169 | 340 | 4 | 24 | 1988 | 6624 | 13644 | 159 | 956 | 37171 | 76566 | 892 | 5367 | 15715 | 32370 | 377 | 2269 | 1618 | 3333 | 39 | 234 | 483 | 995 | 12 | 70 | 298 | 614 | 7 | 43 | 1989 | 3004 | 6114 | 90 | 461 | 15409 | 31367 | 461 | 2365 | 5767 | 11740 | 172 | 885 | 1001 | 2038 | 30 | 154 | 269 | 547 | 8 | 41 | 403 | 820 | 12 | 62 | 1990 | 6750 | 11816 | 236 | 920 | 22244 | 38934 | 776 | 3033 | 9485 | 16602 | 331 | 1293 | 1312 | 2297 | 46 | 179 | 193 | 337 | 7 | 26 | 338 | 591 | 12 | 46 | 1991 | 5650 | 9281 | 193 | 750 | 21005 | 34499 | 718 | 2788 | 8793 | 14443 | 301 | 1167 | 799 | 1312 | 27 | 106 | 155 | 254 | 5 | 21 | 47 | 78 | 2 | 6 | 1992 | 11418 | 22836 | 416 | 4095 | 38670 | 77339 | 1408 | 13867 | 14189 | 28377 | 516 | 5088 | 1681 | 3363 | 61 | 603 | 292 | 585 | 11 | 105 | 0 | 0 | 0 | 0 | 0 | 0 | 1993 | 11793 | 22699 | 415 | 1614 | 45610 | 87791 | 1605 | 6242 | 16661 | 32071 | 586 | 2280 | 2574 | 4954 | 91 | 352 | 462 | 890 | 16 | 63 | 422 | 813 | 15 | 58 | 1994 | 13082 | 28738 | 769 | 3268 | 29401 | 64585 | 1729 | 7343 | 9740 | 21395 | 573 | 2433 | 539 | 1183 | 32 | 135 | 64 | 141 | 4 | 16 | 111 | 243 | 7 | 28 | 1995 | 10205 | 24587 | 609 | 2665 | 31439 | 75745 | 1877 | 8211 | 11108 | 26762 | 663 | 2901 | 386 | 931 | 23 | 101 | 233 | 560 | 14 | 61 | 185 | 446 | 11 | 48 | 1996 | 19519 | 43650 | 1439 | 4273 | 52515 | 117438 | 3870 | 11497 | 17384 | 38875 | 1281 | 3806 | 643 | 1438 | 47 | 141 | 151 | 338 | 11 | 33 | 224 | 500 | 16 | 49 | 1997 | 11763 | 21437 | 1226 | 3970 | 24074 | 43872 | 2509 | 8125 | 6468 | 11786 | 674 | 2183 | 235 | 429 | 25 | 79 | 60 | 110 | 6 | 20 | 60 | 110 | 6 | 20 | 1998 | 19617 | 27571 | 1956 | 6992 | 52347 | 73573 | 5219 | 18658 | 11863 | 16673 | 1183 | 4228 | 538 | 756 | 54 | 192 | 249 | 350 | 25 | 89 | 161 | 227 | 16 | 58 | 1999 | 13981 | 20350 | 1286 | 4196 | 62141 | 90450 | 5717 | 18651 | 10474 | 15245 | 964 | 3143 | 405 | 589 | 37 | 122 | 69 | 100 | 6 | 21 | 151 | 220 | 14 | 45 | 2000 | 19313 | 26033 | 1466 | 3728 | 75551 | 50618 | 2850 | 7248 | 12414 | 16734 | 942 | 2396 | 1128 | 1520 | 86 | 218 | 159 | 214 | 12 | 31 | 106 | 143 | 8 | 20 | 2001 | 11754 | 15383 | 907 | 2104 | 39901 | 52218 | 3080 | 7143 | 10007 | 13095 | 773 | 1791 | 296 | 387 | 23 | 53 | 53 | 69 | 4 | 9 | 20 | 26 | 2 | 4 | 2002 | 10500 | 15736 | 684 | 2006 | 34310 | 51418 | 2234 | 6556 | 3870 | 5799 | 252 | 739 | 241 | 361 | 16 | 46 | 0 | 0 | 0 | 72 | 108 | 5 | 14 | 2003 | 21615 | 26166 | 1092 | 3485 | 74615 | 90328 | 3768 | 12033 | 6583 | 7970 | 332 | 1062 | 458 | 555 | 23 | 74 | 104 | 126 | 5 | 17 | 52 | 63 | 3 | 8 | 2004 | 7992 | 12452 | 396 | 1686 | 49598 | 77280 | 2455 | 10464 | 8385 | 13065 | 415 | 1769 | 180 | 281 | 9 | 38 | 0 | 0 | 0 | 41 | 64 | 2 | 9 | 2005 | 6421 | 18899 | 487 | 2678 | 36753 | 108180 | 2790 | 15329 | 5309 | 15627 | 403 | 2214 | 114 | 336 | 9 | 48 | 0 | 0 | 0 | 26 | 76 | 2 | 11 | 2006 | 10757 | 17194 | 1251 | 3239 | 42745 | 68322 | 4971 | 12872 | 8571 | 13700 | 997 | 2581 | 69 | 110 | 8 | 21 | 0 | 0 | 0 | 172 | 275 | 20 | 52 | 2007 | 10422 | 21117 | 1182 | 3828 | 36934 | 74834 | 4188 | 13567 | 8734 | 17696 | 990 | 3208 | 78 | 157 | 9 | 28 | 129 | 262 | 15 | 47 | 17 | 35 | 2 | 6 | 2008 | 13901 | 23285 | 1062 | 3396 | 63476 | 106328 | 4851 | 15808 | 11459 | 19195 | 876 | 2800 | 330 | 552 | 25 | 81 | 84 | 141 | 6 | 21 | 196 | 329 | 15 | 48 | 2009 | 13313 | 24903 | 787 | 5088 | 59555 | 111403 | 3518 | 22760 | 10610 | 19847 | 627 | 4055 | 485 | 908 | 29 | 185 | 0 | 0 | 0 | 135 | 252 | 8 | 52 | 2010 | 21058 | 26262 | 1610 | 4596 | 79694 | 99392 | 6094 | 17393 | 23093 | 28801 | 1766 | 5040 | 997 | 1243 | 76 | 218 | 211 | 263 | 16 | 46 | 110 | 137 | 8 | 24 | 2011 | 15720 | 26791 | 1308 | 6277 | 60515 | 103137 | 5033 | 24165 | 14418 | 24574 | 1199 | 5758 | 850 | 1448 | 71 | 339 | 100 | 170 | 8 | 40 | 272 | 464 | 23 | 109 | 2012 | 23561 | 33459 | 1662 | 4417 | 72540 | 103017 | 5117 | 13600 | 16241 | 23065 | 1146 | 3045 | 827 | 1174 | 58 | 155 | 112 | 159 | 8 | 21 | 408 | 580 | 29 | 77 | 2013 | 9283 | 13679 | 518 | 4063 | 53415 | 78712 | 2983 | 23377 | 17957 | 26461 | 1003 | 7859 | 860 | 1267 | 48 | 376 | 291 | 429 | 16 | 127 | 126 | 185 | 7 | 55 | 2014 | 19093 | 34681 | 1502 | 4436 | 53945 | 97988 | 4244 | 12533 | 12416 | 22554 | 977 | 2885 | 573 | 1041 | 45 | 133 | 353 | 642 | 28 | 82 | 172 | 312 | 14 | 40 |
| Bugs label SFA3Sm_L SFA3Sm_H SFA3Lg_L SFA3Lg_H SFA4Sm_L SFA4Sm_H SFA4Lg_L SFA4Lg_H SFA5Sm_L SFA5Sm_H SFA5Lg_L SFA5Lg_H SFA6Sm_L SFA6Sm_H SFA6Lg_L SFA6Lg_H SFA7Sm_L SFA7Sm_H SFA7Lg_L SFA7Lg_H SFA8Sm_L SFA8Sm_H SFA8Lg_L SFA8Lg_H | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |

Appendix 4.vi. (Continued). Input data for returns of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

| Year | Salmon Fishing Area 9 | | | | | | Salmon Fishing Area 10 | | | | | | Salmon Fishing Area 11 | | | | | | Salmon Fishing Area 12 | | | | | | Salmon Fishing Area 13 | | | | | | Salmon Fishing Area 14A | | | | | |
|------|-----------------------|-------|--|--------------|------|--|------------------------|-------|--|--------------|------|--|------------------------|-------|--|--------------|------|--|------------------------|-------|--|--------------|------|--|------------------------|-------|--|--------------|-------|--|-------------------------|-------|--|------|-------|--|
| | Small salmon | | | Large salmon | | | Small salmon | | | Large salmon | | | Small salmon | | | Large salmon | | | Small salmon | | | Large salmon | | | Small salmon | | | Large salmon | | | | | | | | |
| | Returns | | | Returns | | | Returns | | | Returns | | | Returns | | | Returns | | | Returns | | | Returns | | | Returns | | | Returns | | | | | | | | |
| | Min | Max | | Min | Max | | Min | Max | | Min | Max | | Min | Max | | Min | Max | | Min | Max | | Min | Max | | Min | Max | | Min | Max | | | | | | | |
| 1970 | 6310 | 12620 | | 373 | 1780 | | 2003 | 4007 | | 119 | 565 | | 16760 | 33520 | | 992 | 4727 | | 2497 | 4993 | | 148 | 704 | | 25942 | 38282 | | 3251 | 5060 | | 14817 | 29633 | | 365 | 2571 | |
| 1971 | 5400 | 10800 | | 320 | 1523 | | 3093 | 6187 | | 183 | 872 | | 13533 | 27067 | | 801 | 3817 | | 1513 | 3027 | | 90 | 427 | | 26011 | 40151 | | 2678 | 4750 | | 12523 | 25047 | | 308 | 2173 | |
| 1972 | 3797 | 7593 | | 225 | 1071 | | 1890 | 3780 | | 112 | 533 | | 16350 | 32700 | | 968 | 4611 | | 3093 | 6187 | | 183 | 872 | | 23526 | 37589 | | 3107 | 5169 | | 8057 | 16113 | | 198 | 1398 | |
| 1973 | 7200 | 14400 | | 426 | 2031 | | 5950 | 11900 | | 352 | 1678 | | 16187 | 32373 | | 958 | 4565 | | 2153 | 4307 | | 127 | 607 | | 27287 | 40227 | | 3303 | 5200 | | 17607 | 35213 | | 433 | 3055 | |
| 1974 | 4980 | 9960 | | 574 | 1149 | | 4040 | 8080 | | 466 | 932 | | 14920 | 29840 | | 1720 | 3441 | | 2193 | 4387 | | 253 | 506 | | 19274 | 28824 | | 2913 | 4257 | | 10400 | 20800 | | 902 | 1805 | |
| 1975 | 6240 | 12480 | | 880 | 1760 | | 1423 | 2847 | | 201 | 401 | | 15003 | 30007 | | 2116 | 4232 | | 1700 | 3400 | | 240 | 479 | | 33671 | 54424 | | 4497 | 7424 | | 16060 | 32120 | | 507 | 1015 | |
| 1976 | 5410 | 10820 | | 651 | 1303 | | 2433 | 4867 | | 293 | 586 | | 13880 | 27760 | | 1671 | 3343 | | 990 | 1980 | | 119 | 238 | | 29382 | 46902 | | 3378 | 5488 | | 24603 | 49207 | | 1437 | 2874 | |
| 1977 | 3600 | 7200 | | 340 | 680 | | 3657 | 7313 | | 346 | 691 | | 13653 | 27307 | | 1290 | 2581 | | 1860 | 3720 | | 176 | 352 | | 17610 | 25240 | | 2877 | 3598 | | 19023 | 38047 | | 666 | 1331 | |
| 1978 | 4343 | 8687 | | 257 | 514 | | 5317 | 10633 | | 315 | 629 | | 13320 | 26640 | | 788 | 1576 | | 1220 | 2440 | | 72 | 144 | | 17807 | 27681 | | 4716 | 5289 | | 10803 | 21607 | | 266 | 532 | |
| 1979 | 5680 | 11360 | | 326 | 651 | | 2830 | 5660 | | 162 | 324 | | 14433 | 22867 | | 655 | 1311 | | 2443 | 4887 | | 140 | 280 | | 20372 | 31829 | | 1183 | 1862 | | 21927 | 43853 | | 233 | 467 | |
| 1980 | 7930 | 15860 | | 356 | 712 | | 5080 | 10160 | | 228 | 456 | | 16897 | 33793 | | 759 | 1518 | | 2733 | 5467 | | 123 | 246 | | 26538 | 38871 | | 5236 | 5913 | | 12477 | 24953 | | 694 | 1388 | |
| 1981 | 6207 | 12413 | | 825 | 1650 | | 4390 | 8780 | | 584 | 1167 | | 23540 | 47080 | | 3129 | 6258 | | 3533 | 7057 | | 470 | 939 | | 31359 | 45989 | | 5148 | 7452 | | 19607 | 39213 | | 1090 | 2180 | |
| 1982 | 6083 | 12167 | | 147 | 293 | | 4187 | 8373 | | 101 | 202 | | 24460 | 48920 | | 590 | 1180 | | 5183 | 10367 | | 125 | 250 | | 31628 | 46698 | | 3442 | 3831 | | 15877 | 31753 | | 3094 | 6189 | |
| 1983 | 7677 | 15353 | | 352 | 705 | | 3800 | 7600 | | 174 | 349 | | 15897 | 31793 | | 730 | 1460 | | 2223 | 4447 | | 102 | 204 | | 20828 | 31701 | | 4465 | 5100 | | 12667 | 25333 | | 1704 | 3407 | |
| 1984 | 7989 | 17023 | | 125 | 1221 | | 5141 | 10955 | | 81 | 785 | | 24767 | 52774 | | 389 | 3784 | | 6782 | 14451 | | 106 | 1036 | | 26184 | 37852 | | 2296 | 3710 | | 16962 | 36143 | | 266 | 2591 | |
| 1985 | 6375 | 13198 | | 96 | 912 | | 4831 | 10000 | | 73 | 691 | | 21213 | 43914 | | 320 | 3034 | | 3996 | 8273 | | 60 | 572 | | 16028 | 25505 | | 1375 | 2508 | | 13209 | 27345 | | 199 | 1890 | |
| 1986 | 8411 | 17555 | | 208 | 1231 | | 5619 | 11727 | | 139 | 822 | | 20300 | 42368 | | 501 | 2970 | | 3433 | 7166 | | 85 | 502 | | 22881 | 36916 | | 2079 | 3649 | | 18411 | 38426 | | 455 | 2694 | |
| 1987 | 3416 | 6865 | | 88 | 489 | | 1690 | 3397 | | 44 | 242 | | 15087 | 30317 | | 391 | 2162 | | 3274 | 6580 | | 85 | 469 | | 19629 | 32325 | | 1546 | 3022 | | 18203 | 36580 | | 471 | 2608 | |
| 1988 | 5179 | 10668 | | 124 | 748 | | 4308 | 8873 | | 103 | 622 | | 18985 | 39106 | | 456 | 2741 | | 5330 | 10979 | | 128 | 770 | | 26162 | 43480 | | 1950 | 3917 | | 23580 | 48570 | | 566 | 3405 | |
| 1989 | 5352 | 10895 | | 160 | 821 | | 3655 | 7440 | | 109 | 561 | | 12047 | 24524 | | 360 | 1849 | | 2279 | 4640 | | 68 | 350 | | 10154 | 16156 | | 849 | 1565 | | 13036 | 26537 | | 390 | 2001 | |
| 1990 | 7332 | 12834 | | 256 | 1000 | | 3281 | 5743 | | 115 | 447 | | 17470 | 30578 | | 610 | 2382 | | 3363 | 5887 | | 117 | 459 | | 21518 | 31183 | | 1778 | 3084 | | 19843 | 34732 | | 693 | 2706 | |
| 1991 | 2404 | 3949 | | 82 | 319 | | 988 | 1622 | | 34 | 131 | | 7956 | 13068 | | 272 | 1056 | | 2765 | 4542 | | 95 | 367 | | 16225 | 20945 | | 1709 | 2433 | | 15307 | 25141 | | 523 | 2031 | |
| 1992 | 5044 | 10088 | | 184 | 1809 | | 1791 | 3582 | | 65 | 642 | | 16615 | 33231 | | 605 | 5958 | | 4671 | 9342 | | 170 | 1675 | | 25990 | 44119 | | 3087 | 8928 | | 34927 | 69854 | | 1271 | 12525 | |
| 1993 | 11402 | 21948 | | 401 | 1560 | | 5578 | 10736 | | 196 | 763 | | 24574 | 47301 | | 865 | 3363 | | 5936 | 11426 | | 209 | 812 | | 27523 | 46889 | | 2618 | 4746 | | 13116 | 59893 | | 1095 | 4258 | |
| 1994 | 3007 | 6607 | | 177 | 751 | | 2544 | 5588 | | 150 | 635 | | 7649 | 16803 | | 450 | 1910 | | 2761 | 6066 | | 162 | 690 | | 22103 | 37166 | | 3476 | 5879 | | 13321 | 29263 | | 783 | 3327 | |
| 1995 | 5321 | 12821 | | 318 | 1390 | | 4371 | 10532 | | 261 | 1142 | | 10757 | 25916 | | 642 | 2809 | | 2294 | 5527 | | 137 | 599 | | 27022 | 49781 | | 1843 | 5096 | | 20840 | 50209 | | 1244 | 5443 | |
| 1996 | 6015 | 13450 | | 443 | 1317 | | 8245 | 18438 | | 608 | 1805 | | 18938 | 42350 | | 1396 | 4146 | | 5025 | 11238 | | 370 | 1100 | | 36576 | 67672 | | 3479 | 7132 | | 32761 | 73263 | | 2415 | 7172 | |
| 1997 | 3636 | 6627 | | 379 | 1227 | | 5071 | 9242 | | 528 | 1712 | | 16648 | 30339 | | 1735 | 5619 | | 4556 | 8303 | | 475 | 1538 | | 31402 | 46494 | | 4240 | 8521 | | 25241 | 45998 | | 2630 | 8519 | |
| 1998 | 4694 | 6597 | | 468 | 1673 | | 7821 | 10992 | | 780 | 2788 | | 8467 | 11900 | | 844 | 3018 | | 2360 | 3318 | | 235 | 841 | | 21816 | 27955 | | 3194 | 7080 | | 23995 | 33724 | | 2392 | 8552 | |
| 1999 | 4015 | 5844 | | 369 | 1205 | | 5113 | 7443 | | 470 | 1535 | | 9643 | 14036 | | 887 | 2894 | | 1139 | 1658 | | 105 | 342 | | 32407 | 40858 | | 3878 | 7739 | | 26960 | 39241 | | 2480 | 8091 | |
| 2000 | 7850 | 10582 | | 596 | 1515 | | 7639 | 10297 | | 580 | 1475 | | 17560 | 23266 | | 1310 | 3332 | | 2634 | 3551 | | 200 | 509 | | 54330 | 67784 | | 5519 | 10048 | | 36819 | 49632 | | 2795 | 7107 | |
| 2001 | 2043 | 2674 | | 158 | 366 | | 2924 | 3826 | | 226 | 523 | | 9396 | 12296 | | 725 | 1682 | | 2201 | 2880 | | 170 | 394 | | 37393 | 45761 | | 3749 | 6510 | | 20775 | 27188 | | 1604 | 3719 | |
| 2002 | 1917 | 2873 | | 125 | 366 | | 3713 | 5565 | | 242 | 709 | | 9011 | 13505 | | 587 | 1722 | | 2321 | 3478 | | 151 | 443 | | 34070 | 46011 | | 3452 | 6469 | | 26558 | 39801 | | 1729 | 5075 | |
| 2003 | 2229 | 2699 | | 113 | 359 | | 3771 | 4565 | | 190 | 608 | | 14208 | 17201 | | 718 | 2291 | | 5917 | 7163 | | 299 | 954 | | 50367 | 57997 | | 4421 | 8434 | | 40802 | 49395 | | 2061 | 6579 | |
| 2004 | 1926 | 3001 | | 95 | 406 | | 3697 | 5760 | | 183 | 780 | | 13762 | 21443 | | 681 | 2903 | | 3131 | 4879 | | 155 | 661 | | 49924 | 66549 | | 4308 | 9118 | | 30057 | 46833 | | 1488 | 6341 | |
| 2005 | 1948 | 5734 | | 148 | 813 | | 2779 | 8180 | | 211 | 1159 | | 6260 | 18425 | | 475 | 2611 | | 2686 | 7905 | | 204 | 1120 | | 40658 | 88340 | | 4595 | 12966 | | 17340 | 51040 | | 1316 | 7232 | |
| 2006 | 4355 | 6960 | | 506 | 1311 | | 5344 | 8542 | | 622 | 1609 | | 11033 | 17634 | | 1283 | 3322 | | 3460 | 5530 | | 402 | 1042 | | 53311 | 74546 | | 8499 | 15058 | | 28081 | 44883 | | 3266 | 8456 | |
| 2007 | 2377 | 4817 | | 270 | 873 | | 3497 | 7086 | | 397 | 1285 | | 5650 | 11449 | | 641 | 2076 | | 2808 | 5689 | | 318 | 1031 | | 33808 | 59140 | | 4691 | 10959 | | 19966 | 40454 | | 2264 | 7334 | |
| 2008 | 3944 | 6606 | | 301 | 963 | | 4786 | 8016 | | 366 | 1169 | | 11136 | 18654 | | 851 | 2721 | | 2610 | 4373 | | 200 | 638 | | 51933 | 75122 | | 3901 | 9668 | | 25802 | 43220 | | 1972 | 6304 | |
| 2009 | 3445 | 6443 | | 203 | 1316 | | 5137 | 9608 | | 303 | 963 | | 7536 | 14097 | | 445 | 2880 | | 1746 | 3266 | | 103 | 667 | | 36368 | 55458 | | 3722 | 10806 | | 21146 | 39555 | | 1249 | 8081 | |
| 2010 | 6597 | 8227 | | 504 | 1440 | | </ | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |

Appendix 4.vii. Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

| Year | Salmon Fishing Area 3 | | | | Salmon Fishing Area 4 | | | | Salmon Fishing Area 5 | | | | Salmon Fishing Area 6 | | | | Salmon Fishing Area 7 | | | | Salmon Fishing Area 8 | | | | |
|-------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|-----------------------|-------|--------------|------|-----------------------|--------|--------------|-------|-----------------------|-------|--------------|------|-----------------------|------|--------------|-----|-----------------------|-----|--------------|-----|-----------------------|-----|--------------|-----|-----|
| | Small salmon | | Large salmon | | Small salmon | | Large salmon | | Small salmon | | Large salmon | | Small salmon | | Large salmon | | Small salmon | | Large salmon | | Small salmon | | Large salmon | | |
| | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | |
| Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max |
| 1970 | 1829 | 4443 | 154 | 736 | 11314 | 27478 | 910 | 4512 | 5194 | 12614 | 404 | 2058 | 196 | 476 | 14 | 76 | 47 | 113 | 3 | 18 | 43 | 105 | 0 | 13 | |
| 1971 | 1731 | 4205 | 135 | 687 | 8827 | 21437 | 688 | 3499 | 3920 | 9520 | 293 | 1541 | 128 | 312 | 10 | 51 | 93 | 227 | 8 | 38 | 58 | 142 | 0 | 15 | |
| 1972 | 1162 | 2822 | 98 | 468 | 8036 | 19516 | 655 | 3214 | 4422 | 10738 | 354 | 1762 | 278 | 674 | 23 | 112 | 142 | 346 | 12 | 57 | 65 | 159 | 6 | 26 | |
| 1973 | 2772 | 6732 | 232 | 1115 | 15657 | 38023 | 1275 | 6259 | 4928 | 11968 | 405 | 1974 | 583 | 1417 | 49 | 235 | 306 | 742 | 26 | 123 | 219 | 533 | 15 | 84 | |
| 1974 | 1958 | 4754 | 318 | 641 | 12537 | 30447 | 1983 | 4049 | 3820 | 9276 | 608 | 1237 | 707 | 1717 | 115 | 232 | 310 | 754 | 49 | 100 | 119 | 289 | 20 | 39 | |
| 1975 | 2583 | 6273 | 520 | 1041 | 13867 | 33677 | 2628 | 5421 | 4639 | 11265 | 912 | 1846 | 219 | 533 | 43 | 87 | 93 | 227 | 19 | 38 | 203 | 493 | 41 | 82 | |
| 1976 | 2210 | 5366 | 379 | 759 | 15594 | 37870 | 2495 | 5177 | 4429 | 10755 | 697 | 1459 | 576 | 1400 | 97 | 196 | 70 | 170 | 12 | 24 | 187 | 453 | 32 | 64 | |
| 1977 | 3570 | 8670 | 478 | 960 | 19591 | 47577 | 1559 | 4204 | 10771 | 26157 | 1410 | 2864 | 936 | 2272 | 107 | 234 | 182 | 442 | 24 | 48 | 189 | 459 | 26 | 51 | |
| 1978 | 1769 | 4295 | 149 | 298 | 20473 | 49719 | 1229 | 2959 | 6669 | 16195 | 536 | 1100 | 691 | 1677 | 51 | 110 | 231 | 567 | 19 | 38 | 103 | 249 | 9 | 17 | |
| 1979 | 4760 | 11560 | 390 | 779 | 18727 | 45481 | 1206 | 2740 | 3106 | 7542 | 234 | 489 | 569 | 1383 | 45 | 91 | 292 | 708 | 24 | 48 | 233 | 567 | 19 | 38 | |
| 1980 | 4067 | 9877 | 224 | 485 | 21966 | 53346 | 903 | 2312 | 6305 | 15311 | 376 | 780 | 747 | 1813 | 34 | 82 | 238 | 578 | 14 | 30 | 280 | 680 | 18 | 36 | |
| 1981 | 5502 | 13362 | 1042 | 2087 | 31584 | 76704 | 5637 | 11635 | 8139 | 19765 | 1511 | 3056 | 1412 | 3428 | 239 | 507 | 287 | 697 | 53 | 107 | 180 | 436 | 34 | 68 | |
| 1982 | 6146 | 14926 | 124 | 336 | 23270 | 56514 | 544 | 1346 | 5677 | 13787 | 143 | 338 | 672 | 1632 | 6 | 29 | 362 | 878 | 2 | 15 | 198 | 482 | 0 | 5 | |
| 1983 | 3773 | 9163 | 245 | 493 | 20893 | 50739 | 1073 | 2443 | 5500 | 13356 | 191 | 551 | 691 | 1677 | 35 | 81 | 324 | 788 | 0 | 9 | 96 | 232 | 1 | 8 | |
| 1984 | 2531 | 6525 | 55 | 540 | 25033 | 64536 | 533 | 5322 | 6835 | 17620 | 149 | 1456 | 789 | 2034 | 12 | 163 | 243 | 626 | 1 | 48 | 200 | 515 | 4 | 43 | |
| 1985 | 3462 | 8569 | 72 | 683 | 32218 | 79741 | 671 | 6352 | 9208 | 22791 | 192 | 1816 | 1134 | 2806 | 24 | 224 | 296 | 733 | 6 | 58 | 272 | 674 | 6 | 54 | |
| 1986 | 2054 | 5126 | 70 | 413 | 24722 | 61700 | 840 | 4977 | 10782 | 26910 | 366 | 2170 | 1184 | 2955 | 40 | 238 | 271 | 677 | 9 | 55 | 367 | 916 | 12 | 74 | |
| 1987 | 1655 | 3895 | 57 | 318 | 16032 | 37722 | 556 | 3078 | 4892 | 11511 | 170 | 939 | 403 | 948 | 14 | 77 | 82 | 194 | 3 | 16 | 126 | 297 | 4 | 24 | |
| 1988 | 4868 | 11888 | 159 | 956 | 27317 | 66712 | 892 | 5367 | 11549 | 28204 | 377 | 2269 | 1189 | 2904 | 39 | 234 | 355 | 867 | 12 | 70 | 219 | 535 | 7 | 43 | |
| 1989 | 2266 | 5376 | 90 | 461 | 11623 | 27581 | 461 | 2365 | 4350 | 10323 | 172 | 885 | 755 | 1792 | 30 | 154 | 203 | 481 | 8 | 41 | 304 | 721 | 12 | 62 | |
| 1990 | 5032 | 10098 | 236 | 920 | 16583 | 33273 | 776 | 3033 | 7071 | 14188 | 331 | 1293 | 978 | 1963 | 46 | 179 | 144 | 288 | 7 | 26 | 252 | 505 | 12 | 46 | |
| 1991 | 4334 | 7965 | 193 | 750 | 16113 | 29607 | 718 | 2788 | 6745 | 12395 | 301 | 1167 | 613 | 1126 | 27 | 106 | 119 | 218 | 5 | 21 | 36 | 67 | 2 | 6 | |
| 1992 | 9844 | 21262 | 415 | 4094 | 33228 | 71898 | 1407 | 13866 | 12175 | 26363 | 516 | 5088 | 1450 | 3132 | 61 | 603 | 252 | 545 | 11 | 105 | 0 | 0 | 0 | 0 | |
| 1993 | 10054 | 20961 | 400 | 1599 | 39162 | 81344 | 1590 | 6226 | 14370 | 29779 | 576 | 2270 | 2243 | 4623 | 90 | 351 | 404 | 831 | 16 | 63 | 369 | 760 | 15 | 58 | |
| 1994 | 9146 | 24802 | 749 | 3247 | 20576 | 55760 | 1644 | 7259 | 6855 | 18510 | 560 | 2420 | 381 | 1026 | 30 | 133 | 46 | 122 | 4 | 16 | 79 | 212 | 6 | 27 | |
| 1995 | 7409 | 21791 | 580 | 2636 | 22872 | 67179 | 1801 | 8135 | 8122 | 23776 | 642 | 2880 | 287 | 831 | 23 | 100 | 173 | 501 | 14 | 60 | 135 | 397 | 11 | 48 | |
| 1996 | 15729 | 39860 | 1412 | 4247 | 42346 | 107268 | 3757 | 11383 | 14095 | 35586 | 1263 | 3787 | 522 | 1317 | 46 | 139 | 124 | 311 | 11 | 33 | 180 | 457 | 16 | 48 | |
| 1997 | 9422 | 19095 | 1209 | 3954 | 19309 | 39107 | 2467 | 8083 | 5228 | 10547 | 668 | 2177 | 190 | 384 | 24 | 79 | 49 | 99 | 6 | 20 | 48 | 98 | 6 | 20 | |
| 1998 | 16390 | 24345 | 1933 | 6969 | 43559 | 64785 | 5160 | 18599 | 9943 | 14753 | 1155 | 4201 | 455 | 673 | 53 | 191 | 212 | 313 | 25 | 88 | 135 | 201 | 16 | 57 | |
| 1999 | 11804 | 18173 | 1279 | 4189 | 52390 | 80698 | 5650 | 18583 | 8832 | 13603 | 947 | 3126 | 343 | 528 | 37 | 121 | 58 | 90 | 6 | 21 | 119 | 188 | 14 | 45 | |
| 2000 | 17003 | 23723 | 1449 | 3711 | 32879 | 45946 | 2803 | 7201 | 10897 | 15217 | 923 | 2377 | 993 | 1386 | 84 | 217 | 140 | 195 | 12 | 31 | 88 | 125 | 8 | 20 | |
| 2001 | 9861 | 13489 | 892 | 2089 | 33365 | 45682 | 3023 | 7086 | 8344 | 11433 | 767 | 1786 | 250 | 342 | 23 | 53 | 42 | 59 | 4 | 9 | 17 | 23 | 2 | 4 | |
| 2002 | 8620 | 13856 | 671 | 1994 | 28099 | 45208 | 2175 | 6498 | 3194 | 5124 | 250 | 737 | 199 | 319 | 15 | 45 | 0 | 0 | 0 | 0 | 55 | 91 | 5 | 14 | |
| 2003 | 19386 | 23938 | 1085 | 3478 | 67026 | 82739 | 3738 | 12001 | 5926 | 7312 | 331 | 1060 | 412 | 508 | 23 | 74 | 94 | 116 | 5 | 17 | 47 | 58 | 3 | 8 | |
| 2004 | 6942 | 11402 | 390 | 1680 | 43104 | 70785 | 2430 | 10438 | 7307 | 11987 | 412 | 1766 | 158 | 259 | 9 | 38 | 0 | 0 | 0 | 0 | 35 | 58 | 2 | 9 | |
| 2005 | 5056 | 17534 | 473 | 2664 | 28896 | 100323 | 2695 | 15235 | 4200 | 14518 | 394 | 2205 | 92 | 314 | 8 | 47 | 0 | 0 | 0 | 0 | 18 | 69 | 2 | 11 | |
| 2006 | 9402 | 15839 | 1228 | 3216 | 37156 | 62732 | 4925 | 12825 | 7495 | 12623 | 969 | 2554 | 61 | 102 | 8 | 20 | 0 | 0 | 0 | 0 | 141 | 244 | 20 | 52 | |
| 2007 | 9147 | 19842 | 1171 | 3818 | 32243 | 70143 | 4122 | 13501 | 7641 | 16603 | 978 | 3196 | 68 | 148 | 8 | 28 | 112 | 245 | 12 | 45 | 15 | 33 | 2 | 6 | |
| 2008 | 11799 | 21183 | 1045 | 3379 | 53591 | 96443 | 4745 | 15402 | 9669 | 17405 | 867 | 2791 | 274 | 497 | 22 | 78 | 69 | 125 | 4 | 18 | 159 | 292 | 15 | 48 | |
| 2009 | 11205 | 22795 | 779 | 5080 | 49881 | 101928 | 3491 | 22732 | 8828 | 18065 | 622 | 4049 | 412 | 834 | 28 | 185 | 0 | 0 | 0 | 0 | 111 | 228 | 7 | 51 | |
| 2010 | 18364 | 23569 | 1595 | 4581 | 69075 | 88772 | 6006 | 17304 | 20114 | 25822 | 1754 | 5028 | 874 | 1120 | 76 | 217 | 183 | 235 | 16 | 46 | 93 | 120 | 8 | 24 | |
| 2011 | 13193 | 24264 | 1291 | 6261 | 50806 | 93428 | 4789 | 23920 | 12075 | 22230 | 1176 | 5734 | 716 | 1314 | 70 | 339 | 83 | 153 | 8 | 39 | 220 | 412 | 22 | 108 | |
| 2012 | 21149 | 31048 | 1639 | 4394 | 64959 | 95436 | 5046 | 13528 | 14554 | 21377 | 1140 | 3039 | 738 | 1086 | 57 | 154 | 100 | 147 | 8 | 21 | 361 | 533 | 25 | 73 | |
| 2013 | 7822 | 12219 | 495 | 4039 | 44838 | 70136 | 2889 | 23284 | 15027 | 23531 | 976 | 7833 | 721 | 1128 | 44 | 373 | 241 | 379 | 16 | 127 | 102 | 162 | 7 | 55 | |
| 2014 | 16973 | 32561 | 1487 | 4420 | 47987 | 92030 | 4190 | 12478 | 11000 | 21137 | 961 | 2869 | 510 | 978 | 43 | 131 | 306 | 595 | 26 | 80 | 152 | 293 | 14 | 40 | |
| SFA3SSm L1 SFA3SSm H1 SFA3SLq SFA3SLq SFA4SSm SFA4SSm SFA4SLq SFA4SLq SFA5SSm SFA5SSm SFA5SLq SFA5SLq SFA6SSm SFA6SSm SFA6SLq SFA6SLq SFA7SSm SFA7SSm SFA7SLq SFA7SLq SFA8SSm SFA8SSm SFA8SLq SFA8SLq | | | | | | | | | | | | | | | | | | | | | | | | | |

Appendix 4.vii. (Continued). Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

| Year | Salmon Fishing Area 9 | | | | Salmon Fishing Area 10 | | | | Salmon Fishing Area 11 | | | | Salmon Fishing Area 12 | | | | Salmon Fishing Area 13 | | | | Salmon Fishing Area 14A | | | |
|------|-----------------------|-------|--------------|------|------------------------|-------|--------------|------|------------------------|-------|--------------|------|------------------------|-------|--------------|------|------------------------|-------|--------------|-------|-------------------------|-------|--------------|-------|
| | Small salmon | | Large salmon | | Small salmon | | Large salmon | | Small salmon | | Large salmon | | Small salmon | | Large salmon | | Small salmon | | Large salmon | | Small salmon | | Large salmon | |
| | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | | Spawners | |
| | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max |
| 1970 | 4417 | 10727 | 361 | 1768 | 1402 | 3406 | 112 | 558 | 11732 | 28492 | 918 | 4653 | 1748 | 4244 | 69 | 625 | 16203 | 28543 | 1608 | 3417 | 10372 | 25188 | 134 | 2340 |
| 1971 | 3780 | 9180 | 301 | 1504 | 2165 | 5259 | 166 | 855 | 9473 | 23007 | 736 | 3752 | 1059 | 2573 | 74 | 411 | 16489 | 30629 | 1633 | 3705 | 8766 | 21290 | 0 | 1850 |
| 1972 | 2658 | 6454 | 217 | 1063 | 1323 | 3213 | 108 | 529 | 11445 | 27795 | 882 | 4525 | 2165 | 5259 | 163 | 852 | 15125 | 29188 | 2004 | 4066 | 5640 | 13696 | 83 | 1283 |
| 1973 | 5040 | 12240 | 406 | 2011 | 4165 | 10115 | 310 | 1636 | 11331 | 27517 | 923 | 4530 | 1507 | 3661 | 102 | 582 | 17019 | 29959 | 1911 | 3808 | 12325 | 29931 | 91 | 2713 |
| 1974 | 3486 | 8466 | 565 | 1140 | 2828 | 6868 | 452 | 918 | 10444 | 25364 | 1682 | 3403 | 1535 | 3729 | 240 | 493 | 12085 | 21635 | 1997 | 3341 | 7280 | 17680 | 789 | 1692 |
| 1975 | 4368 | 10608 | 874 | 1754 | 996 | 2420 | 192 | 392 | 10502 | 25506 | 2076 | 4192 | 1190 | 2890 | 220 | 459 | 21668 | 42421 | 3611 | 6538 | 11242 | 27302 | 417 | 925 |
| 1976 | 3787 | 9197 | 639 | 1291 | 1703 | 4137 | 283 | 576 | 9716 | 23596 | 1629 | 3301 | 693 | 1683 | 114 | 233 | 18999 | 36519 | 2752 | 4862 | 17222 | 41826 | 1337 | 2774 |
| 1977 | 2520 | 6120 | 331 | 671 | 2560 | 6216 | 341 | 686 | 9557 | 23211 | 1272 | 2563 | 1302 | 3162 | 128 | 304 | 10898 | 18528 | 1828 | 2549 | 13316 | 32340 | 194 | 859 |
| 1978 | 3040 | 7384 | 240 | 497 | 3722 | 9038 | 273 | 587 | 9324 | 22644 | 770 | 1558 | 854 | 2074 | 52 | 124 | 12518 | 22392 | 3861 | 4434 | 7562 | 18366 | 194 | 460 |
| 1979 | 3976 | 9656 | 311 | 636 | 1981 | 4811 | 154 | 316 | 8003 | 19437 | 648 | 1304 | 1710 | 4154 | 130 | 270 | 14363 | 25820 | 1070 | 1749 | 15349 | 37275 | 174 | 408 |
| 1980 | 5551 | 13481 | 295 | 651 | 3556 | 8636 | 201 | 429 | 11828 | 28724 | 715 | 1474 | 1913 | 4647 | 94 | 217 | 18625 | 30958 | 4243 | 4920 | 8734 | 21210 | 514 | 1208 |
| 1981 | 4345 | 10551 | 773 | 1598 | 3073 | 7463 | 555 | 1138 | 16478 | 40018 | 3088 | 6217 | 2473 | 6007 | 453 | 922 | 22059 | 36689 | 4485 | 6789 | 13725 | 33331 | 953 | 2043 |
| 1982 | 4258 | 10342 | 114 | 260 | 2931 | 7117 | 91 | 192 | 17122 | 41582 | 537 | 1127 | 3628 | 8812 | 110 | 235 | 22062 | 37132 | 2847 | 3236 | 11114 | 26990 | 2987 | 6082 |
| 1983 | 5374 | 13050 | 281 | 634 | 2660 | 6460 | 95 | 270 | 11128 | 27024 | 703 | 1433 | 1556 | 3780 | 94 | 196 | 14491 | 25364 | 3855 | 4490 | 8867 | 21533 | 1635 | 3338 |
| 1984 | 5725 | 14759 | 120 | 1216 | 3684 | 9498 | 79 | 783 | 17748 | 45755 | 374 | 3769 | 4860 | 12529 | 38 | 968 | 18413 | 30081 | 1987 | 3401 | 12155 | 31336 | 179 | 2504 |
| 1985 | 4625 | 11448 | 96 | 912 | 3505 | 8674 | 73 | 691 | 15390 | 38091 | 320 | 3034 | 2899 | 7176 | 57 | 569 | 10726 | 20203 | 1349 | 2482 | 9583 | 23719 | 197 | 1887 |
| 1986 | 6113 | 15257 | 208 | 1231 | 4084 | 10192 | 139 | 822 | 14754 | 36822 | 501 | 2970 | 2495 | 6228 | 81 | 499 | 15535 | 29570 | 2013 | 3583 | 13381 | 33396 | 445 | 2683 |
| 1987 | 2549 | 5998 | 88 | 489 | 1261 | 2968 | 44 | 242 | 11258 | 26488 | 391 | 2162 | 2443 | 5749 | 82 | 466 | 13611 | 26307 | 1512 | 2988 | 13583 | 31960 | 467 | 2604 |
| 1988 | 3806 | 9295 | 124 | 748 | 3166 | 7731 | 103 | 622 | 13952 | 34073 | 456 | 2741 | 3917 | 9566 | 126 | 767 | 17945 | 35263 | 1909 | 3877 | 17329 | 42319 | 549 | 3388 |
| 1989 | 4037 | 9580 | 160 | 821 | 2757 | 6542 | 109 | 561 | 9087 | 21564 | 360 | 1849 | 1719 | 4080 | 67 | 349 | 6980 | 12982 | 836 | 1552 | 9833 | 23334 | 385 | 1996 |
| 1990 | 5466 | 10968 | 256 | 1000 | 2446 | 4908 | 115 | 447 | 13024 | 26132 | 610 | 2382 | 2507 | 5031 | 114 | 456 | 14866 | 24531 | 1744 | 3051 | 14793 | 29682 | 679 | 2692 |
| 1991 | 1844 | 3389 | 82 | 319 | 758 | 1392 | 34 | 131 | 6103 | 11215 | 272 | 1056 | 2121 | 3898 | 93 | 365 | 11037 | 15757 | 1689 | 2413 | 11742 | 21576 | 512 | 2020 |
| 1992 | 4334 | 9378 | 183 | 1809 | 1496 | 3287 | 65 | 642 | 14239 | 30854 | 605 | 3958 | 3985 | 8657 | 162 | 1667 | 20506 | 38635 | 2992 | 8833 | 30096 | 65023 | 1234 | 12488 |
| 1993 | 9956 | 20502 | 400 | 1559 | 4809 | 9967 | 194 | 761 | 21423 | 44150 | 861 | 3350 | 5176 | 10666 | 207 | 810 | 22341 | 41708 | 2544 | 4673 | 97010 | 55787 | 1058 | 4221 |
| 1994 | 2124 | 5723 | 172 | 746 | 1804 | 4848 | 144 | 630 | 5295 | 14449 | 430 | 1891 | 1949 | 5253 | 154 | 681 | 15381 | 30444 | 3207 | 5611 | 9385 | 25327 | 742 | 3286 |
| 1995 | 3887 | 11386 | 304 | 1376 | 3218 | 9378 | 253 | 1133 | 7770 | 22930 | 625 | 2792 | 1689 | 4922 | 130 | 592 | 20570 | 43329 | 1607 | 4860 | 15218 | 44587 | 1187 | 5385 |
| 1996 | 4868 | 12304 | 431 | 1304 | 6687 | 16880 | 592 | 1789 | 15226 | 38638 | 1362 | 4113 | 4082 | 10295 | 358 | 1088 | 29056 | 60152 | 3199 | 6852 | 26584 | 67085 | 2357 | 7115 |
| 1997 | 2927 | 5918 | 372 | 1221 | 4086 | 8257 | 519 | 1702 | 13304 | 26995 | 1718 | 5602 | 3655 | 7401 | 464 | 1527 | 25508 | 40599 | 3985 | 8266 | 20359 | 41117 | 2578 | 8467 |
| 1998 | 3937 | 5840 | 458 | 1663 | 6606 | 9777 | 771 | 2779 | 7024 | 10457 | 836 | 3009 | 1968 | 2925 | 225 | 831 | 18279 | 24417 | 3031 | 6918 | 19992 | 29721 | 2347 | 8507 |
| 1999 | 3401 | 5230 | 359 | 1195 | 4313 | 6642 | 455 | 1520 | 8086 | 12478 | 881 | 2889 | 958 | 1477 | 102 | 339 | 28647 | 37098 | 3760 | 7621 | 22659 | 34941 | 2402 | 8013 |
| 2000 | 6913 | 9645 | 581 | 1501 | 6664 | 9322 | 534 | 1429 | 14895 | 20901 | 1288 | 3310 | 2291 | 3208 | 195 | 504 | 48055 | 61508 | 5250 | 9779 | 32314 | 45127 | 2731 | 7044 |
| 2001 | 1709 | 2339 | 151 | 359 | 2436 | 3338 | 215 | 513 | 7804 | 10704 | 714 | 1671 | 1818 | 2497 | 162 | 386 | 31037 | 39405 | 3536 | 6297 | 17331 | 23744 | 1559 | 3674 |
| 2002 | 1562 | 2518 | 118 | 360 | 3049 | 4901 | 231 | 699 | 7547 | 11840 | 581 | 1716 | 1896 | 3053 | 147 | 439 | 28083 | 40025 | 3313 | 6330 | 21764 | 35007 | 1668 | 5013 |
| 2003 | 1985 | 2454 | 109 | 355 | 3368 | 4162 | 185 | 603 | 12701 | 15693 | 703 | 2276 | 5282 | 6528 | 288 | 943 | 45027 | 52657 | 4206 | 8218 | 36597 | 45189 | 1988 | 6506 |
| 2004 | 1674 | 2749 | 91 | 402 | 3210 | 5273 | 177 | 774 | 11863 | 19544 | 660 | 2882 | 2704 | 4452 | 149 | 655 | 43889 | 60513 | 4074 | 8883 | 26116 | 42892 | 1429 | 6282 |
| 2005 | 1478 | 5264 | 130 | 794 | 2171 | 7572 | 194 | 1142 | 3827 | 16992 | 456 | 2591 | 2062 | 7282 | 191 | 1107 | 33349 | 81031 | 4320 | 12691 | 13676 | 47376 | 1246 | 7163 |
| 2006 | 3791 | 6397 | 498 | 1302 | 4627 | 7824 | 602 | 1590 | 9554 | 16155 | 1271 | 3310 | 2986 | 5056 | 392 | 1032 | 46296 | 67532 | 8247 | 14807 | 24532 | 41334 | 3210 | 8400 |
| 2007 | 2063 | 4502 | 263 | 867 | 3047 | 6636 | 387 | 1275 | 4907 | 10706 | 636 | 2071 | 2442 | 5323 | 314 | 1027 | 29402 | 54734 | 4511 | 10780 | 17446 | 37934 | 2222 | 7293 |
| 2008 | 3285 | 5948 | 293 | 955 | 3971 | 7202 | 351 | 1154 | 9314 | 16832 | 841 | 2711 | 2178 | 3940 | 193 | 631 | 43277 | 66465 | 3580 | 9346 | 21887 | 39305 | 1915 | 6246 |
| 2009 | 2835 | 5834 | 198 | 1311 | 4193 | 8665 | 298 | 1957 | 6203 | 12763 | 442 | 2877 | 1450 | 2970 | 100 | 664 | 31106 | 50196 | 3526 | 10610 | 17820 | 36229 | 1200 | 8032 |
| 2010 | 5703 | 7334 | 496 | 1432 | 7062 | 9081 | 616 | 1774 | 6859 | 8842 | 604 | 1742 | 2606 | 3347 | 226 | 651 | 49703 | 58889 | 5478 | 10747 | 27468 | 35298 | 2358 | 6848 |
| 2011 | 4364 | 8077 | 433 | 2099 | 7477 | 13826 | 716 | 3566 | 5696 | 10554 | 564 | 2744 | 2074 | 3827 | 203 | 990 | 33849 | 62267 | 3160 | 15915 | 20249 | 37231 | 1953 | 9575 |
| 2012 | 5898 | 8720 | 471 | 1256 | 7488 | 11027 | 581 | 1566 | 5993 | 8819 | 468 | 1255 | 2348 | 3450 | 184 | 490 | 44778 | 65820 | 3395 | 9251 | 31467 | 46268 | 2451 | 6571 |
| 2013 | 3973 | 6228 | 254 | 2071 | 6681 | 10498 | 424 | 3602 | 6130 | 9621 | 398 | 3213 | 1701 | 2668 | 104 | 885 | 28314 | 44299 | 3301 | 13638 | 11746 | 18341 | 734 | 6050 |
| 2014 | 2915 | 5613 | 258 | 766 | 4075 | 7849 | 355 | 1065 | 4131 | 7967 | 365 | 1087 | 2175 | 4179 | 189 | 566 | 39890 | 76610 | 5211 | 15130 | 26917 | 51738 | 2340 | 7011 |

H[] SFA9SSm SFA9SSm SFA9SLg SFA9SLg SFA10SSr SFA10SSr SFA10SLg SFA10SLg SFA11SSr SFA11SSr SFA11SLg SFA11SLg SFA12SSr SFA12SSr SFA12SLg SFA12SLg SFA13SSr SFA13SSr SFA13SLg SFA13SLg SFA14ASSm SFA14ASSm SFA14ASLg SFA14ASLg

Appendix 4.viii. Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

| Year | Salmon Fishing Area 3 | | | | | Salmon Fishing Area 4 | | | | | Salmon Fishing Area 5 | | | | | Salmon Fishing Area 6 | | | | | Salmon Fishing Area 7 | | | | | Salmon Fishing Area 8 | | | | |
|-------------|-----------------------|------------|----------|----------|----------|-----------------------|----------|----------|----------|----------|-----------------------|----------|----------|----------|----------|-----------------------|----------|----------|----------|----------|-----------------------|----------|----------|----------|---------|-----------------------|-----|-----|-----|--|
| | 2SW | | 2SW | | Returns | 2SW | | 2SW | | Returns | 2SW | | 2SW | | Returns | 2SW | | 2SW | | Returns | 2SW | | 2SW | | Returns | 2SW | | | | |
| | Min | Max | Min | Max | | Min | Max | Min | Max | | Min | Max | Min | Max | | Min | Max | Min | Max | | Min | Max | Min | Max | | Min | Max | Min | Max | |
| 1970 | 15 | 147 | 15 | 147 | 96 | 912 | 91 | 902 | 44 | 419 | 40 | 412 | 2 | 16 | 1 | 15 | 0 | 4 | 0 | 4 | 0 | 3 | 0 | 3 | | | | | | |
| 1971 | 15 | 140 | 14 | 137 | 75 | 711 | 69 | 700 | 33 | 316 | 29 | 308 | 1 | 10 | 1 | 10 | 1 | 8 | 1 | 8 | 0 | 5 | 0 | 3 | | | | | | |
| 1972 | 10 | 94 | 10 | 94 | 68 | 648 | 66 | 643 | 37 | 356 | 35 | 352 | 2 | 22 | 2 | 22 | 1 | 11 | 1 | 11 | 1 | 5 | 1 | 5 | | | | | | |
| 1973 | 23 | 223 | 23 | 223 | 132 | 1262 | 127 | 1252 | 42 | 397 | 40 | 395 | 5 | 47 | 5 | 47 | 3 | 25 | 3 | 25 | 2 | 18 | 1 | 17 | | | | | | |
| 1974 | 32 | 129 | 32 | 128 | 207 | 826 | 198 | 810 | 63 | 252 | 61 | 247 | 12 | 47 | 12 | 46 | 5 | 20 | 5 | 20 | 2 | 8 | 2 | 8 | | | | | | |
| 1975 | 52 | 208 | 52 | 208 | 279 | 1117 | 263 | 1084 | 93 | 374 | 91 | 369 | 4 | 18 | 4 | 17 | 2 | 8 | 2 | 8 | 4 | 16 | 4 | 16 | | | | | | |
| 1976 | 38 | 152 | 38 | 152 | 268 | 1073 | 249 | 1035 | 76 | 305 | 70 | 292 | 10 | 40 | 10 | 39 | 1 | 5 | 1 | 5 | 3 | 13 | 3 | 13 | | | | | | |
| 1977 | 48 | 193 | 48 | 192 | 264 | 1058 | 156 | 841 | 145 | 582 | 141 | 573 | 13 | 51 | 11 | 47 | 2 | 10 | 2 | 10 | 3 | 10 | 3 | 10 | | | | | | |
| 1978 | 15 | 60 | 15 | 60 | 173 | 692 | 123 | 592 | 56 | 226 | 54 | 220 | 6 | 23 | 5 | 22 | 2 | 8 | 2 | 8 | 1 | 3 | 1 | 3 | | | | | | |
| 1979 | 39 | 156 | 39 | 156 | 153 | 613 | 121 | 548 | 25 | 102 | 23 | 98 | 5 | 19 | 4 | 18 | 2 | 10 | 2 | 10 | 2 | 8 | 2 | 8 | | | | | | |
| 1980 | 26 | 104 | 22 | 97 | 141 | 564 | 90 | 462 | 40 | 162 | 38 | 156 | 5 | 19 | 3 | 16 | 2 | 6 | 1 | 6 | 2 | 7 | 2 | 7 | | | | | | |
| 1981 | 104 | 418 | 104 | 417 | 600 | 2399 | 564 | 2327 | 155 | 618 | 151 | 611 | 27 | 107 | 24 | 101 | 5 | 22 | 5 | 21 | 3 | 14 | 3 | 14 | | | | | | |
| 1982 | 21 | 85 | 12 | 67 | 80 | 321 | 54 | 269 | 20 | 78 | 14 | 68 | 2 | 9 | 1 | 6 | 1 | 5 | 0 | 3 | 1 | 3 | 0 | 1 | | | | | | |
| 1983 | 25 | 99 | 25 | 99 | 137 | 548 | 107 | 489 | 36 | 144 | 19 | 110 | 5 | 18 | 4 | 16 | 2 | 9 | 0 | 2 | 1 | 3 | 0 | 2 | | | | | | |
| 1984 | 6 | 108 | 6 | 108 | 55 | 1067 | 53 | 1064 | 15 | 291 | 15 | 291 | 2 | 34 | 1 | 33 | 1 | 10 | 0 | 10 | 0 | 9 | 0 | 9 | | | | | | |
| 1985 | 7 | 137 | 7 | 137 | 67 | 1270 | 67 | 1270 | 19 | 363 | 19 | 363 | 2 | 45 | 2 | 45 | 1 | 12 | 1 | 12 | 1 | 11 | 1 | 11 | | | | | | |
| 1986 | 7 | 83 | 7 | 83 | 84 | 995 | 84 | 995 | 37 | 434 | 37 | 434 | 4 | 48 | 4 | 48 | 1 | 11 | 1 | 11 | 1 | 15 | 1 | 15 | | | | | | |
| 1987 | 6 | 64 | 6 | 64 | 56 | 616 | 56 | 616 | 17 | 188 | 17 | 188 | 1 | 15 | 1 | 15 | 0 | 3 | 0 | 3 | 0 | 5 | 0 | 5 | | | | | | |
| 1988 | 16 | 191 | 16 | 191 | 89 | 1073 | 89 | 1073 | 38 | 454 | 38 | 454 | 4 | 47 | 4 | 47 | 1 | 14 | 1 | 14 | 1 | 9 | 1 | 9 | | | | | | |
| 1989 | 9 | 92 | 9 | 92 | 46 | 473 | 46 | 473 | 17 | 177 | 17 | 177 | 3 | 31 | 3 | 31 | 1 | 8 | 1 | 8 | 1 | 12 | 1 | 12 | | | | | | |
| 1990 | 24 | 184 | 24 | 184 | 78 | 607 | 78 | 607 | 33 | 259 | 33 | 259 | 5 | 36 | 5 | 36 | 1 | 5 | 1 | 5 | 1 | 9 | 1 | 9 | | | | | | |
| 1991 | 19 | 150 | 19 | 150 | 72 | 558 | 72 | 558 | 30 | 233 | 30 | 233 | 3 | 21 | 3 | 21 | 1 | 4 | 1 | 4 | 0 | 1 | 0 | 1 | | | | | | |
| 1992 | 42 | 819 | 42 | 819 | 141 | 2773 | 141 | 2773 | 52 | 1018 | 52 | 1018 | 6 | 121 | 6 | 121 | 1 | 21 | 1 | 21 | 0 | 0 | 0 | 0 | | | | | | |
| 1993 | 42 | 323 | 40 | 320 | 161 | 1248 | 159 | 1245 | 59 | 456 | 58 | 454 | 9 | 70 | 9 | 70 | 2 | 13 | 2 | 13 | 1 | 12 | 1 | 12 | | | | | | |
| 1994 | 46 | 457 | 45 | 455 | 104 | 1028 | 99 | 1016 | 34 | 341 | 34 | 339 | 2 | 19 | 2 | 19 | 0 | 2 | 0 | 2 | 0 | 4 | 0 | 4 | | | | | | |
| 1995 | 37 | 373 | 35 | 369 | 113 | 1150 | 108 | 1139 | 40 | 406 | 39 | 403 | 1 | 14 | 1 | 14 | 1 | 9 | 1 | 8 | 1 | 7 | 1 | 7 | | | | | | |
| 1996 | 86 | 598 | 85 | 595 | 232 | 1610 | 225 | 1594 | 77 | 533 | 76 | 530 | 3 | 20 | 3 | 19 | 1 | 5 | 1 | 5 | 1 | 7 | 1 | 7 | | | | | | |
| 1997 | 74 | 556 | 73 | 554 | 151 | 1138 | 148 | 1132 | 40 | 306 | 40 | 305 | 1 | 11 | 1 | 11 | 0 | 3 | 0 | 3 | 0 | 3 | 0 | 3 | | | | | | |
| 1998 | 117 | 979 | 116 | 976 | 313 | 2612 | 310 | 2604 | 71 | 592 | 69 | 588 | 3 | 27 | 3 | 27 | 1 | 12 | 1 | 12 | 1 | 8 | 1 | 8 | | | | | | |
| 1999 | 77 | 587 | 77 | 586 | 343 | 2611 | 339 | 2602 | 58 | 440 | 57 | 438 | 2 | 17 | 2 | 17 | 0 | 3 | 0 | 3 | 1 | 6 | 1 | 6 | | | | | | |
| 2000 | 88 | 522 | 87 | 520 | 171 | 1015 | 168 | 1008 | 57 | 335 | 55 | 333 | 5 | 30 | 5 | 30 | 1 | 4 | 1 | 4 | 0 | 3 | 0 | 3 | | | | | | |
| 2001 | 39 | 196 | 38 | 194 | 132 | 664 | 130 | 659 | 33 | 167 | 33 | 166 | 1 | 5 | 1 | 5 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | | | | | | |
| 2002 | 29 | 187 | 29 | 185 | 96 | 610 | 94 | 604 | 11 | 69 | 11 | 69 | 1 | 4 | 1 | 4 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | | | | | | |
| 2003 | 47 | 324 | 47 | 323 | 162 | 1119 | 161 | 1116 | 14 | 99 | 14 | 99 | 1 | 7 | 1 | 7 | 0 | 2 | 0 | 2 | 0 | 1 | 0 | 1 | | | | | | |
| 2004 | 17 | 157 | 17 | 156 | 106 | 973 | 104 | 971 | 18 | 165 | 18 | 164 | 0 | 4 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | | | | | | |
| 2005 | 21 | 249 | 20 | 248 | 120 | 1426 | 116 | 1417 | 17 | 206 | 17 | 205 | 0 | 4 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | | | | | | |
| 2006 | 54 | 301 | 53 | 299 | 214 | 1197 | 212 | 1193 | 43 | 240 | 42 | 237 | 0 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 5 | 1 | 5 | | | | | | |
| 2007 | 51 | 356 | 50 | 355 | 180 | 1262 | 177 | 1256 | 43 | 298 | 42 | 297 | 0 | 3 | 0 | 3 | 1 | 4 | 1 | 4 | 0 | 1 | 0 | 1 | | | | | | |
| 2008 | 46 | 316 | 45 | 314 | 209 | 1442 | 204 | 1432 | 38 | 260 | 37 | 260 | 1 | 7 | 1 | 7 | 0 | 2 | 0 | 2 | 1 | 4 | 1 | 4 | | | | | | |
| 2009 | 34 | 473 | 33 | 472 | 151 | 2117 | 150 | 2114 | 27 | 377 | 27 | 377 | 1 | 17 | 1 | 17 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 5 | | | | | | |
| 2010 | 69 | 427 | 69 | 426 | 262 | 1618 | 258 | 1609 | 76 | 469 | 75 | 468 | 3 | 20 | 3 | 20 | 1 | 4 | 1 | 4 | 0 | 2 | 0 | 2 | | | | | | |
| 2011 | 56 | 584 | 56 | 582 | 216 | 2247 | 206 | 2225 | 52 | 535 | 51 | 533 | 3 | 32 | 3 | 31 | 0 | 4 | 0 | 4 | 1 | 10 | 1 | 10 | | | | | | |
| 2012 | 71 | 411 | 70 | 409 | 220 | 1265 | 217 | 1258 | 49 | 283 | 49 | 283 | 3 | 14 | 2 | 14 | 0 | 2 | 0 | 2 | 1 | 7 | 1 | 7 | | | | | | |
| 2013 | 22 | 378 | 21 | 376 | 128 | 2174 | 124 | 2165 | 43 | 731 | 42 | 728 | 2 | 35 | 2 | 35 | 1 | 12 | 1 | 12 | 0 | 5 | 0 | 5 | | | | | | |
| 2014 | 65 | 413 | 64 | 411 | 182 | 1166 | 180 | 1160 | 42 | 268 | 41 | 267 | 2 | 12 | 2 | 12 | 1 | 8 | 1 | 7 | 1 | 4 | 1 | 4 | | | | | | |
| Bugs labels | SFA3R2_L[] | SFA3R2_H[] | SFA3S2_L | SFA3S2_L | SFA4R2_L | SFA4R2_L | SFA4S2_L | SFA4S2_L | SFA5R2_L | SFA5R2_L | SFA5S2_L | SFA5S2_L | SFA6R2_L | SFA6R2_L | SFA6S2_L | SFA6S2_L | SFA7R2_L | SFA7R2_L | SFA7S2_L | SFA7S2_L | SFA8R2_L | SFA8R2_L | SFA8S2_L | SFA8S2_L | | | | | | |

Bugs labels: SFA3R2_L() SFA3R2_H() SFA3S2_L SFA3S2_H SFA4R2_L SFA4R2_H SFA4S2_L SFA4S2_H SFA5R2_L SFA5R2_H SFA5S2_L SFA5S2_H SFA6R2_L SFA6R2_H SFA6S2_L SFA6S2_H SFA7R2_L SFA7R2_H SFA7S2_L SFA7S2_H SFA8R2_L SFA8R2_H SFA8S2_L SFA8S2_H

Appendix 4.viii. (Continued). Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

| Year | Salmon Fishing Area 9 | | | | Salmon Fishing Area 10 | | | | Salmon Fishing Area 11 | | | | Salmon Fishing Area 12 | | | | Salmon Fishing Area 13 | | | | Salmon Fishing Area 14A | | | |
|-------------|-----------------------|----------|----------|----------|------------------------|-----------|-----------|-----------|------------------------|-----------|-----------|-----------|------------------------|-----------|-----------|-----------|------------------------|-----------|-----------|-----------|-------------------------|------------|------------|------------|
| | 2SW | | 2SW | | 2SW | | 2SW | | 2SW | | 2SW | | 2SW | | 2SW | | 2SW | | 2SW | | 2SW | | 2SW | |
| | Returns | Max | Spawners | Max | Returns | Max | Spawners | Max | Returns | Max | Spawners | Max | Returns | Max | Spawners | Max | Returns | Max | Spawners | Max | Returns | Max | Spawners | Max |
| 1970 | 37 | 356 | 36 | 354 | 12 | 113 | 11 | 112 | 99 | 945 | 92 | 931 | 15 | 141 | 7 | 125 | 1300 | 3036 | 643 | 2050 | 36 | 514 | 13 | 468 |
| 1971 | 32 | 305 | 30 | 301 | 18 | 174 | 17 | 171 | 80 | 763 | 74 | 750 | 9 | 85 | 7 | 82 | 1071 | 2850 | 653 | 2223 | 31 | 435 | 0 | 370 |
| 1972 | 22 | 214 | 22 | 213 | 11 | 107 | 11 | 106 | 97 | 922 | 88 | 905 | 18 | 174 | 16 | 170 | 1243 | 3101 | 802 | 2439 | 20 | 280 | 8 | 257 |
| 1973 | 43 | 406 | 41 | 402 | 35 | 336 | 31 | 327 | 96 | 913 | 92 | 906 | 13 | 121 | 10 | 116 | 1321 | 3120 | 764 | 2285 | 43 | 611 | 9 | 543 |
| 1974 | 57 | 230 | 57 | 228 | 47 | 186 | 45 | 184 | 172 | 688 | 168 | 681 | 25 | 101 | 24 | 99 | 1165 | 2554 | 799 | 2005 | 90 | 361 | 79 | 338 |
| 1975 | 88 | 352 | 87 | 351 | 20 | 80 | 19 | 78 | 212 | 846 | 208 | 838 | 24 | 96 | 22 | 92 | 1799 | 4454 | 1445 | 3923 | 51 | 203 | 42 | 185 |
| 1976 | 65 | 261 | 64 | 258 | 29 | 117 | 28 | 115 | 167 | 669 | 163 | 660 | 12 | 48 | 11 | 47 | 1351 | 3293 | 1101 | 2917 | 144 | 575 | 134 | 555 |
| 1977 | 34 | 136 | 33 | 134 | 35 | 138 | 34 | 137 | 129 | 516 | 127 | 513 | 18 | 70 | 13 | 61 | 1151 | 2159 | 731 | 1530 | 67 | 266 | 19 | 172 |
| 1978 | 26 | 103 | 24 | 99 | 31 | 126 | 27 | 117 | 79 | 315 | 77 | 312 | 7 | 29 | 5 | 25 | 1886 | 3173 | 1544 | 2660 | 27 | 106 | 19 | 92 |
| 1979 | 33 | 130 | 31 | 127 | 16 | 65 | 15 | 63 | 66 | 262 | 65 | 261 | 14 | 56 | 13 | 54 | 473 | 1117 | 428 | 1049 | 23 | 93 | 17 | 82 |
| 1980 | 36 | 142 | 30 | 130 | 23 | 91 | 20 | 86 | 76 | 304 | 71 | 295 | 12 | 40 | 9 | 43 | 2094 | 3548 | 1697 | 2952 | 69 | 278 | 51 | 242 |
| 1981 | 83 | 330 | 77 | 320 | 58 | 233 | 55 | 228 | 313 | 1252 | 309 | 1243 | 47 | 185 | 45 | 184 | 2059 | 4471 | 1794 | 4073 | 109 | 436 | 95 | 409 |
| 1982 | 15 | 59 | 11 | 52 | 10 | 40 | 9 | 38 | 59 | 236 | 54 | 225 | 13 | 50 | 11 | 47 | 1377 | 2298 | 1139 | 1941 | 309 | 1238 | 299 | 1216 |
| 1983 | 35 | 141 | 28 | 127 | 17 | 70 | 10 | 54 | 73 | 292 | 70 | 287 | 10 | 41 | 9 | 39 | 1786 | 3060 | 1542 | 2694 | 170 | 681 | 163 | 668 |
| 1984 | 13 | 244 | 12 | 243 | 8 | 157 | 8 | 157 | 39 | 757 | 37 | 754 | 11 | 207 | 4 | 194 | 918 | 2226 | 795 | 2041 | 27 | 518 | 18 | 501 |
| 1985 | 10 | 182 | 10 | 182 | 7 | 138 | 7 | 138 | 32 | 607 | 32 | 607 | 6 | 114 | 6 | 114 | 550 | 1505 | 540 | 1489 | 20 | 378 | 20 | 377 |
| 1986 | 21 | 246 | 21 | 246 | 14 | 164 | 14 | 164 | 50 | 594 | 50 | 594 | 8 | 100 | 8 | 100 | 832 | 2190 | 805 | 2150 | 45 | 539 | 44 | 537 |
| 1987 | 9 | 98 | 9 | 98 | 4 | 48 | 4 | 48 | 39 | 432 | 39 | 432 | 8 | 94 | 8 | 93 | 618 | 1813 | 605 | 1793 | 47 | 522 | 47 | 521 |
| 1988 | 12 | 150 | 12 | 150 | 10 | 124 | 10 | 124 | 46 | 548 | 46 | 548 | 13 | 154 | 13 | 153 | 780 | 2350 | 764 | 2326 | 57 | 681 | 55 | 678 |
| 1989 | 16 | 164 | 16 | 164 | 11 | 112 | 11 | 112 | 36 | 370 | 36 | 370 | 7 | 70 | 7 | 70 | 339 | 939 | 334 | 931 | 39 | 400 | 39 | 399 |
| 1990 | 26 | 200 | 26 | 200 | 11 | 89 | 11 | 89 | 61 | 476 | 61 | 476 | 12 | 92 | 11 | 91 | 711 | 1851 | 698 | 1830 | 69 | 541 | 68 | 538 |
| 1991 | 8 | 64 | 8 | 64 | 3 | 26 | 3 | 26 | 27 | 211 | 27 | 211 | 9 | 73 | 9 | 73 | 684 | 1460 | 676 | 1448 | 52 | 406 | 51 | 404 |
| 1992 | 18 | 362 | 18 | 362 | 7 | 128 | 6 | 128 | 60 | 1192 | 60 | 1192 | 17 | 335 | 16 | 333 | 1235 | 5357 | 1197 | 5300 | 127 | 2505 | 123 | 2498 |
| 1993 | 40 | 312 | 40 | 312 | 20 | 153 | 19 | 152 | 86 | 673 | 86 | 672 | 21 | 162 | 21 | 162 | 1047 | 2848 | 1018 | 2804 | 110 | 852 | 106 | 844 |
| 1994 | 11 | 105 | 10 | 104 | 9 | 89 | 9 | 88 | 27 | 267 | 26 | 265 | 10 | 97 | 9 | 95 | 1390 | 3528 | 1283 | 3366 | 47 | 466 | 44 | 460 |
| 1995 | 19 | 195 | 18 | 193 | 16 | 160 | 15 | 159 | 39 | 393 | 38 | 391 | 8 | 84 | 8 | 83 | 737 | 3058 | 643 | 2916 | 75 | 762 | 71 | 754 |
| 1996 | 27 | 184 | 26 | 183 | 36 | 253 | 35 | 250 | 84 | 580 | 82 | 576 | 22 | 154 | 22 | 152 | 1391 | 4279 | 1280 | 4111 | 145 | 1004 | 141 | 996 |
| 1997 | 23 | 172 | 22 | 171 | 32 | 240 | 31 | 238 | 104 | 787 | 103 | 784 | 28 | 215 | 28 | 214 | 1696 | 5113 | 1594 | 4960 | 158 | 1193 | 155 | 1185 |
| 1998 | 28 | 234 | 27 | 233 | 47 | 390 | 46 | 389 | 51 | 422 | 50 | 421 | 14 | 118 | 13 | 116 | 1278 | 4248 | 1212 | 4151 | 144 | 1197 | 141 | 1191 |
| 1999 | 22 | 169 | 22 | 167 | 28 | 215 | 27 | 213 | 53 | 405 | 53 | 404 | 6 | 48 | 6 | 48 | 1551 | 4643 | 1504 | 4573 | 149 | 1133 | 144 | 1122 |
| 2000 | 36 | 212 | 35 | 210 | 35 | 206 | 32 | 200 | 79 | 466 | 77 | 463 | 12 | 71 | 12 | 71 | 2208 | 6029 | 2100 | 5867 | 168 | 995 | 164 | 986 |
| 2001 | 7 | 34 | 7 | 33 | 10 | 49 | 9 | 48 | 31 | 156 | 31 | 155 | 7 | 37 | 7 | 36 | 697 | 2324 | 658 | 2248 | 69 | 346 | 67 | 342 |
| 2002 | 5 | 34 | 5 | 33 | 10 | 66 | 10 | 65 | 25 | 160 | 25 | 160 | 6 | 41 | 6 | 41 | 642 | 2309 | 616 | 2260 | 74 | 472 | 72 | 466 |
| 2003 | 5 | 33 | 5 | 33 | 8 | 57 | 8 | 56 | 31 | 213 | 30 | 212 | 13 | 89 | 12 | 88 | 822 | 3011 | 782 | 2934 | 89 | 612 | 85 | 605 |
| 2004 | 4 | 38 | 4 | 37 | 8 | 73 | 8 | 72 | 29 | 270 | 28 | 268 | 7 | 61 | 6 | 61 | 801 | 3255 | 758 | 3171 | 64 | 590 | 61 | 584 |
| 2005 | 6 | 76 | 6 | 74 | 9 | 108 | 8 | 106 | 20 | 243 | 20 | 241 | 9 | 104 | 8 | 103 | 855 | 4629 | 804 | 4531 | 57 | 673 | 54 | 666 |
| 2006 | 22 | 122 | 21 | 121 | 27 | 150 | 26 | 148 | 55 | 309 | 55 | 308 | 17 | 97 | 17 | 96 | 1581 | 5376 | 1534 | 5286 | 140 | 786 | 138 | 781 |
| 2007 | 12 | 81 | 11 | 81 | 17 | 119 | 17 | 119 | 28 | 193 | 27 | 193 | 14 | 96 | 13 | 95 | 872 | 3912 | 839 | 3849 | 97 | 682 | 96 | 678 |
| 2008 | 13 | 90 | 13 | 89 | 16 | 109 | 15 | 107 | 37 | 253 | 36 | 252 | 9 | 59 | 8 | 59 | 726 | 3451 | 666 | 3337 | 85 | 586 | 82 | 581 |
| 2009 | 9 | 122 | 9 | 122 | 13 | 183 | 13 | 182 | 19 | 268 | 19 | 268 | 4 | 62 | 4 | 62 | 692 | 3858 | 656 | 3788 | 54 | 752 | 52 | 747 |
| 2010 | 22 | 134 | 21 | 133 | 27 | 166 | 26 | 165 | 26 | 163 | 26 | 162 | 10 | 61 | 10 | 61 | 1078 | 3951 | 1019 | 3837 | 104 | 643 | 101 | 637 |
| 2011 | 19 | 196 | 19 | 195 | 32 | 335 | 31 | 332 | 25 | 256 | 24 | 255 | 9 | 92 | 9 | 92 | 144 | 1498 | 136 | 1480 | 86 | 895 | 84 | 890 |
| 2012 | 20 | 117 | 20 | 117 | 26 | 147 | 25 | 146 | 20 | 117 | 20 | 117 | 8 | 46 | 8 | 46 | 152 | 873 | 146 | 860 | 107 | 614 | 105 | 611 |
| 2013 | 11 | 194 | 11 | 193 | 19 | 328 | 18 | 326 | 18 | 300 | 17 | 299 | 5 | 83 | 4 | 82 | 854 | 3687 | 814 | 3621 | 33 | 567 | 32 | 563 |
| 2014 | 11 | 71 | 11 | 71 | 16 | 100 | 15 | 99 | 16 | 102 | 16 | 101 | 8 | 53 | 8 | 53 | 1235 | 5643 | 1207 | 5598 | 103 | 657 | 101 | 652 |
| Bugs labels | SFA9R2_L | SFA9R2_L | SFA9S2_L | SFA9S2_L | SFA10R2_L | SFA10R2_L | SFA10S2_L | SFA10S2_L | SFA11R2_L | SFA11R2_L | SFA11S2_L | SFA11S2_L | SFA12R2_L | SFA12R2_L | SFA12S2_L | SFA12S2_L | SFA13R2_L | SFA13R2_L | SFA13S2_L | SFA13S2_L | SFA14AR2_L | SFA14AR2_L | SFA14AS2_L | SFA14AS2_L |

Appendix 4.ix. Input data for small salmon returns to Québec by category of data used in the run-reconstruction.

| Year | Small returns Minimum | | | | | | | | Small returns Maximum | | | | | | | |
|-------------|--------------------------|------------|------------|------------|------------|------------|------------|--------------|--------------------------|------------|------------|------------|------------|------------|------------|--------------|
| | C1 | C2 | C3 | C4 | C5 | C6 | FN Harvest | Other rivers | C1 | C2 | C3 | C4 | C5 | C6 | FN Harvest | Other rivers |
| Bugs labels | QCSmC1_L[] | QCSmC2_L[] | QCSmC3_L[] | QCSmC4_L[] | QCSmC5_L[] | QCSmC6_L[] | QCSmFn_L[] | QCSmO_L[] | QCSmC1_H[] | QCSmC2_H[] | QCSmC3_H[] | QCSmC4_H[] | QCSmC5_H[] | QCSmC6_H[] | QCSmFn_H[] | QCSmO_H[] |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 3830 | 5434 | 2955 | 460 | 1670 | 5160 | 267 | 31 | 4085 | 5639 | 6053 | 792 | 2784 | 8599 | 445 | 52 |
| 1985 | 5266 | 2271 | 1767 | 210 | 5449 | 4384 | 267 | 40 | 5869 | 2336 | 3586 | 352 | 9224 | 7307 | 445 | 67 |
| 1986 | 8648 | 5193 | 2396 | 63 | 6719 | 5133 | 267 | 77 | 9471 | 5321 | 4895 | 107 | 11198 | 8555 | 445 | 129 |
| 1987 | 10043 | 4775 | 3852 | 327 | 8396 | 5501 | 267 | 71 | 10869 | 4910 | 7875 | 546 | 13993 | 9168 | 445 | 118 |
| 1988 | 11190 | 5968 | 4404 | 468 | 8440 | 6423 | 267 | 85 | 12244 | 6133 | 8962 | 780 | 14067 | 10705 | 445 | 142 |
| 1989 | 10121 | 4743 | 2924 | 301 | 6744 | 5622 | 267 | 68 | 10910 | 4878 | 5940 | 503 | 11240 | 9369 | 445 | 113 |
| 1990 | 12245 | 7332 | 4377 | 694 | 7096 | 2976 | 377 | 77 | 13279 | 7511 | 8917 | 1158 | 11826 | 4960 | 628 | 129 |
| 1991 | 9554 | 5851 | 3776 | 349 | 5009 | 2001 | 256 | 57 | 10249 | 5987 | 7679 | 584 | 8348 | 3336 | 426 | 95 |
| 1992 | 9188 | 6928 | 4567 | 428 | 5131 | 3462 | 243 | 70 | 9847 | 7144 | 9297 | 715 | 8552 | 5770 | 405 | 117 |
| 1993 | 8143 | 6325 | 3973 | 1029 | 4315 | 1447 | 525 | 55 | 8883 | 6517 | 8075 | 1717 | 7192 | 2412 | 875 | 92 |
| 1994 | 8707 | 5928 | 3840 | 1051 | 4011 | 437 | 408 | 30 | 9442 | 6129 | 7828 | 1753 | 6686 | 729 | 681 | 50 |
| 1995 | 6943 | 3439 | 2697 | 1017 | 3853 | 434 | 184 | 30 | 7538 | 3527 | 5471 | 1696 | 6422 | 723 | 306 | 50 |
| 1996 | 15010 | 1809 | 3600 | 477 | 4666 | 500 | 120 | 5 | 16122 | 1923 | 7370 | 797 | 7816 | 833 | 200 | 8 |
| 1997 | 11491 | 201 | 3457 | 292 | 3529 | 462 | 58 | 563 | 12089 | 242 | 7049 | 487 | 5882 | 770 | 97 | 938 |
| 1998 | 11285 | 1183 | 3578 | 328 | 5121 | 1127 | 58 | 0 | 11849 | 1406 | 7347 | 555 | 8536 | 1878 | 97 | 0 |
| 1999 | 10877 | 708 | 3194 | 1868 | 5401 | 1429 | 0 | 0 | 11556 | 741 | 6536 | 3098 | 9002 | 2382 | 0 | 0 |
| 2000 | 11886 | 429 | 1116 | 602 | 7399 | 633 | 0 | 0 | 12635 | 458 | 2284 | 1004 | 14050 | 1055 | 0 | 0 |
| 2001 | 8050 | 185 | 2632 | 266 | 3225 | 728 | 0 | 0 | 8588 | 228 | 5392 | 443 | 5374 | 1213 | 0 | 0 |
| 2002 | 14599 | 31 | 3189 | 689 | 4333 | 1448 | 0 | 0 | 15494 | 36 | 6530 | 1149 | 7222 | 2414 | 0 | 0 |
| 2003 | 11362 | 0 | 3203 | 721 | 3566 | 1512 | 0 | 0 | 11903 | 0 | 6538 | 1201 | 5944 | 2520 | 0 | 0 |
| 2004 | 13747 | 107 | 6526 | 284 | 4889 | 1639 | 0 | 0 | 14177 | 127 | 13104 | 474 | 8149 | 2731 | 0 | 0 |
| 2005 | 8771 | 0 | 3689 | 794 | 3353 | 1508 | 0 | 0 | 9188 | 0 | 7485 | 1323 | 5588 | 2513 | 0 | 0 |
| 2006 | 12762 | 0 | 3736 | 1800 | 2944 | 1455 | 0 | 0 | 13369 | 0 | 7584 | 2999 | 4907 | 2426 | 0 | 0 |
| 2007 | 8515 | 0 | 3758 | 1710 | 1830 | 1024 | 0 | 0 | 8964 | 0 | 7631 | 2850 | 3051 | 1707 | 0 | 0 |
| 2008 | 16445 | 0 | 5542 | 2266 | 3144 | 1401 | 0 | 0 | 17350 | 0 | 11261 | 3776 | 5240 | 2336 | 0 | 0 |
| 2009 | 8872 | 0 | 3601 | 903 | 1907 | 1056 | 0 | 0 | 9315 | 0 | 7306 | 1505 | 3178 | 1759 | 0 | 0 |
| 2010 | 12889 | 0 | 4801 | 993 | 1675 | 1081 | 0 | 0 | 13538 | 0 | 9746 | 1655 | 2792 | 1802 | 0 | 0 |
| 2011 | 17993 | 0 | 5120 | 1365 | 4441 | 1694 | 0 | 0 | 18899 | 0 | 10386 | 2276 | 7402 | 2824 | 0 | 0 |
| 2012 | 9566 | 0 | 3615 | 584 | 3550 | 1228 | 0 | 0 | 10038 | 0 | 7332 | 973 | 5916 | 2047 | 0 | 0 |
| 2013 | 7219 | 88 | 3185 | 411 | 2466 | 1401 | 0 | 0 | 7574 | 104 | 6461 | 685 | 4111 | 2335 | 0 | 0 |
| 2014 | 9193 | 0 | 3945 | 660 | 2412 | 2341 | 0 | 0 | 9647 | 0 | 8003 | 1100 | 4020 | 3901 | 0 | 0 |

Appendix 4.ix. (Continued). Input data for large salmon returns to Québec by category of data used in the run-reconstruction.

| Year | Large returns Minimum | | | | | | | | Large returns Maximum | | | | | | | |
|-------------|--------------------------|------------|------------|------------|------------|------------|------------|--------------|--------------------------|------------|------------|------------|------------|------------|------------|--------------|
| | C1 | C2 | C3 | C4 | C5 | C6 | FN Harvest | Other rivers | C1 | C2 | C3 | C4 | C5 | C6 | FN Harvest | Other rivers |
| Bugs labels | QCLgC1_L[] | QCLgC2_L[] | QCLgC3_L[] | QCLgC4_L[] | QCLgC5_L[] | QCLgC6_L[] | QCLgFn_L[] | QCLgO_L[] | QCLgC1_H[] | QCLgC2_H[] | QCLgC3_H[] | QCLgC4_H[] | QCLgC5_H[] | QCLgC6_H[] | QCLgFn_H[] | QCLgO_H[] |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 14119 | 9501 | 2922 | 3407 | 3712 | 5071 | 329 | 108 | 15631 | 9788 | 6035 | 6477 | 6187 | 8452 | 548 | 181 |
| 1985 | 14015 | 7028 | 3836 | 345 | 9215 | 3351 | 329 | 76 | 15611 | 7281 | 7809 | 577 | 15827 | 5586 | 548 | 127 |
| 1986 | 18589 | 8598 | 6152 | 35 | 5877 | 4971 | 329 | 89 | 20602 | 8839 | 12596 | 61 | 9795 | 8284 | 548 | 149 |
| 1987 | 17574 | 6715 | 5178 | 273 | 6335 | 3012 | 329 | 82 | 19017 | 6889 | 10575 | 458 | 10558 | 5019 | 548 | 137 |
| 1988 | 21445 | 6432 | 7540 | 346 | 6789 | 4781 | 329 | 98 | 22979 | 6618 | 15336 | 576 | 11315 | 7969 | 548 | 164 |
| 1989 | 20278 | 8503 | 5530 | 278 | 5718 | 4567 | 329 | 106 | 21906 | 8736 | 11252 | 465 | 9531 | 7611 | 548 | 176 |
| 1990 | 17098 | 10803 | 8164 | 1365 | 5179 | 2424 | 442 | 112 | 18222 | 11041 | 16613 | 2276 | 8631 | 4040 | 737 | 187 |
| 1991 | 19112 | 6988 | 7183 | 696 | 3856 | 357 | 242 | 101 | 20443 | 7192 | 14602 | 1161 | 6427 | 595 | 403 | 168 |
| 1992 | 18392 | 7360 | 7930 | 372 | 2687 | 1503 | 461 | 76 | 19578 | 7560 | 16149 | 622 | 4478 | 2505 | 769 | 127 |
| 1993 | 14578 | 10133 | 2866 | 373 | 2649 | 333 | 423 | 52 | 15454 | 11463 | 5849 | 624 | 4414 | 555 | 705 | 87 |
| 1994 | 16538 | 9172 | 2644 | 506 | 2853 | 145 | 427 | 60 | 17594 | 10241 | 5411 | 845 | 4755 | 242 | 712 | 100 |
| 1995 | 21658 | 9598 | 1926 | 813 | 4390 | 154 | 246 | 31 | 22968 | 10936 | 3915 | 1358 | 7317 | 256 | 410 | 52 |
| 1996 | 22679 | 5822 | 3843 | 577 | 2486 | 135 | 113 | 4 | 24117 | 6941 | 7844 | 964 | 4155 | 225 | 189 | 7 |
| 1997 | 18106 | 4221 | 2816 | 333 | 2865 | 138 | 48 | 9 | 19154 | 5154 | 5768 | 553 | 4775 | 229 | 80 | 15 |
| 1998 | 13180 | 4927 | 2861 | 347 | 2790 | 291 | 48 | 0 | 13891 | 5962 | 5907 | 592 | 4649 | 485 | 80 | 0 |
| 1999 | 16912 | 842 | 2554 | 3661 | 3870 | 492 | 0 | 0 | 17700 | 995 | 5232 | 6103 | 6450 | 838 | 0 | 0 |
| 2000 | 14568 | 619 | 3901 | 560 | 6420 | 563 | 0 | 0 | 15300 | 669 | 7947 | 933 | 10700 | 949 | 0 | 0 |
| 2001 | 17837 | 633 | 5320 | 241 | 3988 | 556 | 0 | 0 | 18889 | 879 | 10914 | 402 | 6647 | 926 | 0 | 0 |
| 2002 | 12335 | 8 | 4515 | 339 | 2103 | 345 | 0 | 0 | 13001 | 9 | 9277 | 565 | 3505 | 575 | 0 | 0 |
| 2003 | 21853 | 0 | 5787 | 269 | 4889 | 384 | 0 | 0 | 22893 | 0 | 11779 | 449 | 8148 | 641 | 0 | 0 |
| 2004 | 18369 | 107 | 4870 | 357 | 4432 | 401 | 0 | 0 | 19043 | 126 | 9170 | 595 | 7387 | 668 | 0 | 0 |
| 2005 | 19154 | 0 | 3204 | 734 | 4615 | 351 | 0 | 0 | 20066 | 0 | 6515 | 1223 | 8025 | 585 | 0 | 0 |
| 2006 | 16704 | 0 | 3387 | 901 | 3945 | 403 | 0 | 0 | 17500 | 0 | 6904 | 1502 | 6575 | 672 | 0 | 0 |
| 2007 | 14832 | 0 | 3638 | 1301 | 3171 | 305 | 0 | 0 | 15604 | 0 | 7406 | 2168 | 5285 | 508 | 0 | 0 |
| 2008 | 15216 | 0 | 5187 | 1328 | 5423 | 390 | 0 | 0 | 16002 | 0 | 10595 | 2213 | 9038 | 649 | 0 | 0 |
| 2009 | 18479 | 0 | 3727 | 950 | 4656 | 275 | 0 | 0 | 19412 | 0 | 7589 | 1584 | 7594 | 458 | 0 | 0 |
| 2010 | 21375 | 0 | 4488 | 1047 | 3656 | 338 | 0 | 0 | 22454 | 0 | 9157 | 1744 | 6093 | 564 | 0 | 0 |
| 2011 | 26977 | 0 | 4697 | 1571 | 6007 | 483 | 0 | 0 | 28373 | 0 | 9529 | 2619 | 10011 | 805 | 0 | 0 |
| 2012 | 17918 | 0 | 3665 | 904 | 4488 | 313 | 0 | 0 | 18837 | 0 | 7434 | 1507 | 7481 | 522 | 0 | 0 |
| 2013 | 22026 | 205 | 4171 | 989 | 3938 | 339 | 0 | 0 | 23135 | 242 | 8461 | 1648 | 6563 | 565 | 0 | 0 |
| 2014 | 10954 | 0 | 2400 | 711 | 1593 | 1035 | 0 | 0 | 11504 | 0 | 4869 | 1185 | 2655 | 1725 | 0 | 0 |

Appendix 4.ix. (Continued). Input data for small salmon spawners to Québec by category of data used in the run-reconstruction.

| Year | Small spawners Minimum | | | | | | Small spawners Maximum | | | | | |
|-------------|---------------------------|-----------|-----------|-----------|-----------|-----------|---------------------------|-----------|-----------|-----------|-----------|-----------|
| | C1 | C2 | C3 | C4 | C5 | C6 | C1 | C2 | C3 | C4 | C5 | C6 |
| Bugs labels | QCSSmC1_L | QCSSmC2_L | QCSSmC3_L | QCSSmC4_L | QCSSmC5_L | QCSSmC6_L | QCSSmC1_H | QCSSmC2_H | QCSSmC3_H | QCSSmC4_H | QCSSmC5_H | QCSSmC6_H |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 3061 | 4342 | 1915 | 415 | 1264 | 5160 | 3316 | 4547 | 5013 | 747 | 2378 | 8599 |
| 1985 | 3960 | 1622 | 1025 | 209 | 4241 | 4384 | 4563 | 1687 | 2844 | 351 | 8016 | 7307 |
| 1986 | 6337 | 3827 | 1499 | 63 | 5151 | 5133 | 7160 | 3955 | 3998 | 107 | 9630 | 8555 |
| 1987 | 7493 | 3489 | 2365 | 291 | 6411 | 5501 | 8319 | 3624 | 6388 | 510 | 12008 | 9168 |
| 1988 | 8173 | 4188 | 2738 | 419 | 6432 | 6423 | 9227 | 4353 | 7296 | 731 | 12059 | 10705 |
| 1989 | 7779 | 3810 | 1878 | 273 | 5149 | 5622 | 8568 | 3945 | 4894 | 475 | 9645 | 9369 |
| 1990 | 8735 | 5757 | 2822 | 604 | 5437 | 2976 | 9768 | 5936 | 7362 | 1068 | 10167 | 4960 |
| 1991 | 7247 | 4551 | 2465 | 316 | 3827 | 2001 | 7942 | 4687 | 6368 | 551 | 7166 | 3336 |
| 1992 | 5989 | 4841 | 2937 | 370 | 3957 | 3462 | 6648 | 5057 | 7667 | 657 | 7378 | 5770 |
| 1993 | 4852 | 4311 | 2524 | 747 | 3339 | 1447 | 5592 | 4503 | 6626 | 1435 | 6216 | 2412 |
| 1994 | 5506 | 3996 | 2501 | 894 | 3089 | 437 | 6241 | 4197 | 6489 | 1596 | 5764 | 729 |
| 1995 | 5348 | 2835 | 1760 | 877 | 2956 | 434 | 5943 | 2923 | 4534 | 1556 | 5525 | 723 |
| 1996 | 10636 | 1330 | 2260 | 372 | 3678 | 500 | 11748 | 1444 | 6030 | 692 | 6828 | 833 |
| 1997 | 8238 | 142 | 2250 | 266 | 3074 | 462 | 8836 | 178 | 5842 | 461 | 5426 | 770 |
| 1998 | 7734 | 995 | 2347 | 289 | 4229 | 1124 | 8298 | 1218 | 6116 | 516 | 7643 | 1875 |
| 1999 | 8155 | 509 | 2495 | 1653 | 4581 | 1426 | 8834 | 542 | 5837 | 2883 | 8182 | 2379 |
| 2000 | 8291 | 372 | 693 | 519 | 5900 | 583 | 9040 | 401 | 1861 | 921 | 12551 | 1005 |
| 2001 | 5329 | 143 | 1870 | 263 | 2579 | 658 | 5867 | 186 | 4140 | 440 | 4729 | 1137 |
| 2002 | 9296 | 31 | 2231 | 658 | 3405 | 1448 | 10191 | 36 | 5572 | 1118 | 6294 | 2414 |
| 2003 | 8180 | 0 | 2269 | 661 | 2826 | 1509 | 8721 | 0 | 5604 | 1141 | 5204 | 2517 |
| 2004 | 9030 | 29 | 5574 | 278 | 3962 | 1639 | 9460 | 49 | 12152 | 468 | 7222 | 2731 |
| 2005 | 6339 | 0 | 3025 | 716 | 2709 | 1506 | 6756 | 0 | 6821 | 1245 | 4945 | 2511 |
| 2006 | 8628 | 0 | 3159 | 1691 | 2372 | 1455 | 9235 | 0 | 7007 | 2890 | 4335 | 2426 |
| 2007 | 5768 | 0 | 3226 | 1511 | 1501 | 1024 | 6217 | 0 | 7099 | 2651 | 2722 | 1707 |
| 2008 | 10562 | 0 | 4882 | 1756 | 2522 | 1401 | 11467 | 0 | 10601 | 3266 | 4618 | 2336 |
| 2009 | 6293 | 0 | 3115 | 764 | 1633 | 1056 | 6736 | 0 | 6820 | 1366 | 2904 | 1759 |
| 2010 | 8860 | 0 | 4289 | 914 | 1311 | 1080 | 9509 | 0 | 9234 | 1576 | 2428 | 1801 |
| 2011 | 12143 | 0 | 4496 | 1116 | 3674 | 1688 | 13049 | 0 | 9762 | 2027 | 6635 | 2818 |
| 2012 | 6620 | 0 | 3152 | 472 | 2924 | 1225 | 7092 | 0 | 6869 | 861 | 5290 | 2044 |
| 2013 | 4959 | 88 | 2840 | 328 | 2131 | 1401 | 5314 | 104 | 6116 | 602 | 3776 | 2335 |
| 2014 | 6579 | 0 | 3239 | 544 | 2127 | 2341 | 7033 | 0 | 7297 | 984 | 3735 | 3901 |

Appendix 4.ix. (Continued). Input data for large salmon spawners to Québec by category of data used in the run-reconstruction.

| Year | Large spawners Minimum | | | | | | Large spawners Maximum | | | | | |
|-------------|---------------------------|-----------|-----------|-----------|-----------|-----------|---------------------------|-----------|-----------|-----------|-----------|-----------|
| | C1 | C2 | C3 | C4 | C5 | C6 | C1 | C2 | C3 | C4 | C5 | C6 |
| Bugs labels | QCSLgC1_L | QCSLgC2_L | QCSLgC3_L | QCSLgC4_L | QCSLgC5_L | QCSLgC6_L | QCSLgC1_H | QCSLgC2_H | QCSLgC3_H | QCSLgC4_H | QCSLgC5_H | QCSLgC6_H |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 10421 | 7648 | 1861 | 2357 | 2815 | 5071 | 11933 | 7935 | 4974 | 5427 | 5290 | 8452 |
| 1985 | 9985 | 4991 | 2125 | 340 | 7214 | 3351 | 11581 | 5244 | 6098 | 572 | 13826 | 5586 |
| 1986 | 13659 | 5804 | 3695 | 35 | 4498 | 4971 | 15672 | 6045 | 10139 | 61 | 8416 | 8284 |
| 1987 | 13432 | 4791 | 3025 | 246 | 4830 | 3012 | 14875 | 4965 | 8422 | 431 | 9053 | 5019 |
| 1988 | 15535 | 4258 | 4381 | 312 | 5172 | 4781 | 17069 | 4444 | 12177 | 542 | 9698 | 7969 |
| 1989 | 14645 | 6742 | 3239 | 253 | 4375 | 4567 | 16273 | 6975 | 8961 | 440 | 8188 | 7611 |
| 1990 | 12398 | 8463 | 4557 | 1228 | 3950 | 2424 | 13522 | 8701 | 13006 | 2139 | 7402 | 4040 |
| 1991 | 14061 | 5019 | 3970 | 596 | 2940 | 357 | 15392 | 5223 | 11389 | 1061 | 5511 | 595 |
| 1992 | 12850 | 4819 | 4492 | 325 | 2044 | 1503 | 14036 | 5019 | 12711 | 575 | 3835 | 2505 |
| 1993 | 9848 | 6936 | 1809 | 282 | 2038 | 333 | 10724 | 8266 | 4792 | 533 | 3803 | 555 |
| 1994 | 10468 | 5920 | 1693 | 448 | 2173 | 145 | 11524 | 6989 | 4460 | 787 | 4075 | 242 |
| 1995 | 16562 | 8323 | 1321 | 781 | 3367 | 154 | 17872 | 9661 | 3310 | 1326 | 6294 | 256 |
| 1996 | 16431 | 4417 | 2389 | 394 | 1924 | 135 | 17869 | 5536 | 6390 | 781 | 3593 | 225 |
| 1997 | 13433 | 3393 | 1744 | 308 | 2237 | 138 | 14481 | 4326 | 4696 | 528 | 4147 | 229 |
| 1998 | 10402 | 4429 | 1849 | 302 | 2213 | 290 | 11113 | 5464 | 4895 | 547 | 4073 | 484 |
| 1999 | 14169 | 747 | 1962 | 3100 | 2956 | 491 | 14957 | 900 | 4640 | 5542 | 5536 | 837 |
| 2000 | 11937 | 570 | 8322 | 491 | 5096 | 363 | 12669 | 620 | 7368 | 864 | 9376 | 749 |
| 2001 | 14527 | 505 | 4281 | 239 | 2980 | 348 | 15579 | 751 | 8986 | 400 | 5639 | 717 |
| 2002 | 10843 | 8 | 4071 | 313 | 1500 | 344 | 11509 | 9 | 8833 | 539 | 2902 | 574 |
| 2003 | 18832 | 0 | 5164 | 267 | 3763 | 383 | 19872 | 0 | 11156 | 447 | 7022 | 640 |
| 2004 | 15558 | 107 | 4231 | 355 | 3268 | 401 | 16232 | 126 | 8531 | 593 | 6223 | 668 |
| 2005 | 16485 | 0 | 2901 | 719 | 3556 | 351 | 17397 | 0 | 6212 | 1208 | 6766 | 585 |
| 2006 | 14977 | 0 | 3055 | 872 | 2863 | 403 | 15773 | 0 | 6572 | 1473 | 5493 | 672 |
| 2007 | 12470 | 0 | 3203 | 1287 | 2444 | 303 | 13242 | 0 | 6971 | 2154 | 4558 | 506 |
| 2008 | 13725 | 0 | 4676 | 1266 | 4296 | 390 | 14511 | 0 | 10084 | 2151 | 7911 | 649 |
| 2009 | 16489 | 0 | 3188 | 849 | 3588 | 275 | 17422 | 0 | 7050 | 1483 | 6626 | 458 |
| 2010 | 19170 | 0 | 3926 | 1023 | 3017 | 338 | 20249 | 0 | 8595 | 1720 | 5454 | 564 |
| 2011 | 24130 | 0 | 4180 | 1497 | 4579 | 479 | 25526 | 0 | 9012 | 2545 | 8583 | 801 |
| 2012 | 16098 | 0 | 3221 | 868 | 3685 | 313 | 17017 | 0 | 6990 | 1471 | 6678 | 522 |
| 2013 | 19804 | 205 | 3701 | 920 | 2925 | 338 | 20913 | 242 | 7991 | 1579 | 5550 | 564 |
| 2014 | 10089 | 0 | 2138 | 680 | 1206 | 1035 | 10639 | 0 | 4607 | 1154 | 2268 | 1725 |

Appendix 4.ix. (Continued). Year specific harvest data (1984 to 2009) and returns and spawners data for Québec for years when category splits are not available (1970 to 1983) used in the run-reconstruction.

| Harvests in various fisheries not in the other inputs | | | | | | | These data are specific to the 1970 to 1983 period when detailed returns by river category are not available. | | | | | | | |
|-------------------------------------------------------|--------------|----------|------------|--------------|----------|------------|---------------------------------------------------------------------------------------------------------------|----------|---------------|----------|----------------|-----------|----------------|-----------|
| Year | Small salmon | | | Large salmon | | | Small returns | | Large returns | | Small spawners | | Large spawners | |
| | Sport | FN | Commercial | Sport | FN | Commercial | Min | Max | Min | Max | Min | Max | Min | Max |
| Bugs labels | QCSportSm[] | QCFnSm[] | QCCmSm[] | QCSportLg[] | QCFnLg[] | QCCmLg[] | QCSm_L[] | QCSm_H[] | QCLg_L[] | QCLg_H[] | QCSSm_L[] | QCSSm_H[] | QCSLg_L[] | QCSLg_H[] |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 | 18904 | 28356 | 82680 | 124020 | 11045 | 16568 | 31292 | 46937 |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 14969 | 22453 | 47354 | 71031 | 9338 | 14007 | 16194 | 24292 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 12470 | 18704 | 61773 | 92660 | 8213 | 12320 | 31727 | 47590 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 16585 | 24877 | 68171 | 102256 | 10987 | 16480 | 32279 | 48419 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 16791 | 25186 | 91455 | 137182 | 10067 | 15100 | 39256 | 58884 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 18071 | 27106 | 77664 | 116497 | 11606 | 17409 | 32627 | 48940 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 19959 | 29938 | 77212 | 115818 | 12979 | 19469 | 31032 | 46548 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 18190 | 27285 | 91017 | 136525 | 12004 | 18006 | 44660 | 66990 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 16971 | 25456 | 81953 | 122930 | 11447 | 17170 | 40944 | 61416 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 21683 | 32524 | 45197 | 67796 | 15863 | 23795 | 17543 | 26315 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 29791 | 44686 | 107461 | 161192 | 20817 | 31226 | 48758 | 73137 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 41667 | 62501 | 84428 | 126642 | 30952 | 46428 | 35798 | 53697 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 23699 | 35548 | 74870 | 112305 | 16877 | 25316 | 36290 | 54435 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 17987 | 26981 | 61488 | 92232 | 12030 | 18045 | 23710 | 35565 |
| 1984 | 3492 | 357 | 794 | 8561 | 4530 | 13053 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 4046 | 273 | 2093 | 9883 | 3623 | 16619 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 6266 | 372 | 3707 | 11643 | 4519 | 20889 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 7443 | 366 | 2992 | 9740 | 4466 | 22745 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 8663 | 397 | 4760 | 12980 | 4747 | 19750 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 6080 | 196 | 2615 | 11040 | 2905 | 18175 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 8581 | 108 | 3425 | 12132 | 2900 | 16092 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 6271 | 265 | 3282 | 11194 | 4335 | 16372 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 8263 | 120 | 3849 | 12291 | 4550 | 15851 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 8319 | 7 | 3627 | 9798 | 3976 | 11242 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 7655 | 161 | 3861 | 10932 | 4496 | 10424 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 4187 | 353 | 3915 | 7892 | 6194 | 10038 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 7265 | 72 | 4532 | 9618 | 6113 | 7454 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 5075 | 35 | 3531 | 6771 | 4875 | 7202 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 5867 | 35 | 1068 | 4702 | 4875 | 1038 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 4428 | 710 | 814 | 4407 | 3683 | 471 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 5553 | 821 | 0 | 4297 | 3818 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 4213 | 770 | 0 | 5558 | 3574 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 7206 | 1672 | 0 | 2484 | 3164 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 4898 | 972 | 0 | 4610 | 3541 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 6633 | 1158 | 0 | 4412 | 3958 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 3767 | 909 | 0 | 3973 | 3062 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 5366 | 1117 | 0 | 3032 | 3512 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 3787 | 869 | 0 | 3419 | 2932 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 7604 | 1171 | 0 | 3038 | 2971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 3444 | 1141 | 0 | 3338 | 2752 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 4917 | 1057 | 0 | 3166 | 2362 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 7378 | 1205 | 0 | 4310 | 3216 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 4075 | 1239 | 0 | 2732 | 3023 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 2968 | 1177 | 0 | 3472 | 2895 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 3670 | 1240 | 0 | 1343 | 2908 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix 4.x. Input data for 2SW salmon returns to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

| Year of return to rivers | Returns of 2SW | | | | | | | | | | | | USA Point estimate |
|-----------------------------|----------------|------------|------------|------------|------------|------------|------------|------------|-------------|-------------|------------|------------|--------------------------|
| | SFA 15 | | SFA 16 | | SFA 17 | | SFA 18 | | SFA 19-21 | | SFA 23 | | |
| | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | |
| Winbugs labels | SF15R2_L[] | SF15R2_H[] | SF16R2_L[] | SF16R2_H[] | SF17R2_L[] | SF17R2_H[] | SF18R2_L[] | SF18R2_H[] | SF19_21R2_L | SF19_21R2_H | SF23R2_L[] | SF23R2_H[] | USAR2[] |
| 1970 | 8243 | 10576 | 42901 | 45798 | 31 | 60 | 4744 | 6836 | 5600 | 7447 | 8540 | 12674 | 0 |
| 1971 | 3587 | 4616 | 26038 | 30669 | 29 | 29 | 1891 | 2782 | 4120 | 5215 | 7155 | 10536 | 653 |
| 1972 | 4980 | 9756 | 29092 | 43510 | 402 | 402 | 4693 | 6024 | 5744 | 6993 | 7869 | 11368 | 1383 |
| 1973 | 6211 | 12009 | 26599 | 40492 | 206 | 206 | 4140 | 5481 | 6922 | 8659 | 4205 | 6036 | 1427 |
| 1974 | 7264 | 14570 | 39270 | 60090 | 386 | 386 | 5481 | 6928 | 13138 | 15363 | 10755 | 14988 | 1394 |
| 1975 | 4353 | 7922 | 25889 | 39325 | 345 | 345 | 3452 | 4340 | 12261 | 13797 | 13107 | 18578 | 2331 |
| 1976 | 7293 | 14416 | 20448 | 30758 | 575 | 578 | 2755 | 3674 | 8607 | 10104 | 14274 | 20281 | 1317 |
| 1977 | 9174 | 18077 | 49881 | 73330 | 606 | 606 | 3985 | 5463 | 10872 | 12851 | 16869 | 23995 | 1998 |
| 1978 | 5458 | 10749 | 19504 | 26041 | 0 | 0 | 4585 | 6265 | 8272 | 9779 | 8225 | 11294 | 4208 |
| 1979 | 1472 | 2535 | 6501 | 9306 | 459 | 463 | 1290 | 2014 | 3781 | 4879 | 5165 | 7207 | 1942 |
| 1980 | 7102 | 14045 | 35163 | 48457 | 2 | 5 | 3732 | 5177 | 14094 | 17318 | 19056 | 26865 | 5796 |
| 1981 | 4572 | 7357 | 11144 | 19268 | 40 | 77 | 2490 | 3769 | 8662 | 11471 | 11026 | 15267 | 5601 |
| 1982 | 4314 | 6313 | 21442 | 41643 | 16 | 31 | 4135 | 5901 | 4458 | 5353 | 9782 | 13871 | 6056 |
| 1983 | 3453 | 5280 | 16349 | 28419 | 17 | 32 | 3733 | 5241 | 4134 | 5356 | 9662 | 13836 | 2155 |
| 1984 | 3329 | 6092 | 12216 | 31455 | 13 | 26 | 2391 | 3573 | 1758 | 2854 | 15706 | 22627 | 3222 |
| 1985 | 4805 | 9500 | 14614 | 37625 | 8 | 15 | 921 | 4481 | 6894 | 12124 | 16541 | 23828 | 5529 |
| 1986 | 7831 | 15403 | 21617 | 55640 | 5 | 11 | 2274 | 11479 | 6755 | 11878 | 9891 | 14261 | 6176 |
| 1987 | 4836 | 9123 | 12524 | 32224 | 66 | 128 | 2611 | 10422 | 3748 | 6591 | 6922 | 10043 | 3081 |
| 1988 | 7152 | 13998 | 14384 | 36938 | 96 | 185 | 2533 | 10205 | 4393 | 7735 | 4716 | 6697 | 3286 |
| 1989 | 4390 | 8492 | 9113 | 23385 | 149 | 287 | 2108 | 8600 | 4808 | 8469 | 6560 | 9437 | 3197 |
| 1990 | 4326 | 8369 | 14269 | 36639 | 284 | 545 | 1893 | 7684 | 3591 | 6320 | 5486 | 7918 | 5051 |
| 1991 | 2387 | 4668 | 14685 | 37736 | 188 | 361 | 2350 | 9628 | 2960 | 5213 | 7337 | 10563 | 2647 |
| 1992 | 4002 | 7787 | 21381 | 30728 | 95 | 183 | 2374 | 9577 | 2633 | 4634 | 6878 | 9809 | 2459 |
| 1993 | 1395 | 2684 | 15579 | 60246 | 22 | 43 | 1341 | 5317 | 2542 | 4470 | 4345 | 4820 | 2231 |
| 1994 | 3960 | 7745 | 13652 | 24887 | 169 | 310 | 1981 | 8094 | 1360 | 2396 | 3084 | 3495 | 1346 |
| 1995 | 2713 | 5333 | 25593 | 37215 | 384 | 576 | 1498 | 6160 | 2253 | 3969 | 3439 | 3998 | 1748 |
| 1996 | 3917 | 7754 | 11126 | 19117 | 394 | 591 | 3247 | 13507 | 3000 | 5278 | 4729 | 5397 | 2407 |
| 1997 | 2488 | 4898 | 8545 | 14244 | 387 | 581 | 3421 | 14254 | 1163 | 2045 | 2769 | 3176 | 1611 |
| 1998 | 1687 | 3260 | 6292 | 10783 | 385 | 577 | 2055 | 8560 | 924 | 1270 | 1372 | 1642 | 1526 |
| 1999 | 1780 | 3425 | 7098 | 11206 | 383 | 575 | 1557 | 6596 | 1419 | 1951 | 2375 | 2640 | 1168 |
| 2000 | 2270 | 4410 | 7560 | 11744 | 378 | 566 | 1467 | 6302 | 1078 | 1483 | 988 | 1206 | 533 |
| 2001 | 3779 | 7442 | 14257 | 19289 | 376 | 564 | 1689 | 7251 | 1822 | 2506 | 1938 | 2279 | 788 |
| 2002 | 2335 | 4540 | 5572 | 9079 | 372 | 557 | 1228 | 5307 | 382 | 525 | 483 | 548 | 504 |
| 2003 | 3947 | 7778 | 10991 | 16823 | 371 | 557 | 2380 | 10207 | 1854 | 2548 | 1056 | 1198 | 1192 |
| 2004 | 3005 | 5886 | 10596 | 18488 | 367 | 550 | 2639 | 11397 | 1028 | 1413 | 1335 | 1605 | 1283 |
| 2005 | 3422 | 6725 | 11310 | 19988 | 373 | 560 | 2217 | 9293 | 662 | 906 | 809 | 1012 | 984 |
| 2006 | 2551 | 4973 | 9779 | 17103 | 392 | 587 | 2114 | 9010 | 1263 | 1734 | 922 | 1171 | 1023 |
| 2007 | 4267 | 8422 | 9451 | 15183 | 412 | 618 | 1353 | 6122 | 603 | 825 | 616 | 736 | 954 |
| 2008 | 2848 | 5572 | 5811 | 11066 | 429 | 644 | 2020 | 9357 | 1793 | 2465 | 812 | 1042 | 1764 |
| 2009 | 3948 | 7781 | 10580 | 17076 | 402 | 602 | 1524 | 7251 | 827 | 1135 | 1485 | 1886 | 2069 |
| 2010 | 2978 | 5831 | 7804 | 11581 | 439 | 658 | 2049 | 9574 | 934 | 1277 | 829 | 992 | 1078 |
| 2011 | 7265 | 14445 | 21216 | 48573 | 653 | 980 | 3633 | 16729 | 1489 | 2044 | 2486 | 3259 | 3045 |
| 2012 | 3230 | 6338 | 7987 | 15163 | 653 | 980 | 831 | 4118 | 623 | 849 | 268 | 331 | 879 |
| 2013 | 5324 | 10544 | 7493 | 15865 | 993 | 1487 | 1871 | 8797 | 2075 | 2852 | 420 | 543 | 525 |
| 2014 | 2707 | 5287 | 5607 | 11379 | 713 | 1067 | 1187 | 5996 | 417 | 571 | 172 | 217 | 334 |

Appendix 4.x. (continued). Input data for large salmon returns to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

| Year of return to rivers | Returns of large salmon | | | | | | | | | | | | | | Point estimate |
|-----------------------------|-------------------------|------------|----------------|------------|----------------|------------|----------------|------------|-------------------|-------------------|----------------|------------|---------|------|-------------------|
| | SFA 15 | | SFA 16 | | SFA 17 | | SFA 18 | | SFA 19-21 | | SFA 23 | | USA | | |
| | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | | | |
| Winbugs labels | SF15Lg_L [] | SF15Lg_H[] | SF16Lg_L [] | SF16Lg_H[] | SF17Lg_L [] | SF17Lg_H[] | SF18Lg_L [] | SF18Lg_H[] | SF19_21L g_L[] | SF19_21L g_H[] | SF23Lg_L [] | SF23Lg_H[] | USALg[] | | |
| 1970 | 12681 | 16270 | 46462 | 49599 | 31 | 60 | 6161 | 7858 | 7273 | 9671 | 9691 | 13945 | | 0 | |
| 1971 | 5518 | 7102 | 28365 | 33409 | 29 | 29 | 2456 | 3198 | 5350 | 6773 | 8056 | 11573 | | 653 | |
| 1972 | 8441 | 16536 | 30146 | 45087 | 402 | 402 | 6095 | 6924 | 7460 | 9082 | 8890 | 12536 | | 1383 | |
| 1973 | 8393 | 16229 | 27771 | 42276 | 206 | 206 | 5376 | 6299 | 8049 | 10069 | 4760 | 6638 | | 1427 | |
| 1974 | 9950 | 19959 | 43249 | 66179 | 386 | 386 | 7119 | 7963 | 13138 | 15363 | 12187 | 16444 | | 1394 | |
| 1975 | 5510 | 10028 | 29826 | 45305 | 345 | 345 | 4483 | 4989 | 12261 | 13797 | 14829 | 20351 | | 2331 | |
| 1976 | 9596 | 18969 | 23943 | 36016 | 575 | 578 | 3578 | 4223 | 8873 | 10416 | 16128 | 22175 | | 1317 | |
| 1977 | 11053 | 21779 | 52673 | 77434 | 606 | 606 | 5175 | 6280 | 14119 | 16690 | 19165 | 26183 | | 1998 | |
| 1978 | 7277 | 14332 | 22653 | 30245 | 0 | 0 | 5954 | 7201 | 10471 | 12378 | 9335 | 12342 | | 4208 | |
| 1979 | 2886 | 4971 | 9435 | 13507 | 459 | 463 | 1676 | 2315 | 5180 | 6684 | 5856 | 7903 | | 1942 | |
| 1980 | 8768 | 17340 | 37014 | 51008 | 2 | 5 | 4846 | 5951 | 16388 | 20137 | 21464 | 29480 | | 5796 | |
| 1981 | 9729 | 15652 | 16708 | 28887 | 40 | 77 | 3234 | 4332 | 11706 | 15501 | 12481 | 16743 | | 5601 | |
| 1982 | 7311 | 10700 | 26504 | 51475 | 16 | 31 | 5370 | 6783 | 9485 | 11390 | 11147 | 15303 | | 6056 | |
| 1983 | 5852 | 8950 | 20309 | 35304 | 17 | 32 | 4848 | 6024 | 6562 | 8501 | 10908 | 15235 | | 2155 | |
| 1984 | 4214 | 7711 | 12941 | 33321 | 13 | 26 | 3105 | 4107 | 2408 | 3909 | 17706 | 24992 | | 3222 | |
| 1985 | 7627 | 15080 | 16798 | 43247 | 8 | 15 | 1196 | 5150 | 8512 | 14968 | 18582 | 26289 | | 5529 | |
| 1986 | 10305 | 20267 | 25342 | 65228 | 5 | 11 | 2953 | 3195 | 10722 | 18854 | 11142 | 15761 | | 6176 | |
| 1987 | 7556 | 14255 | 15734 | 40483 | 66 | 128 | 3391 | 11980 | 5950 | 10462 | 7865 | 11116 | | 3081 | |
| 1988 | 9933 | 19441 | 17627 | 45267 | 96 | 185 | 3289 | 11729 | 7321 | 12891 | 5360 | 7312 | | 3286 | |
| 1989 | 7701 | 14898 | 13955 | 35812 | 149 | 287 | 2738 | 9885 | 6969 | 12275 | 7393 | 10380 | | 3197 | |
| 1990 | 6362 | 12307 | 23164 | 59479 | 284 | 545 | 2458 | 8832 | 6191 | 10897 | 6235 | 8710 | | 5051 | |
| 1991 | 4773 | 9335 | 24273 | 62373 | 188 | 361 | 3052 | 11066 | 4112 | 7240 | 8312 | 11659 | | 2647 | |
| 1992 | 7411 | 14420 | 34573 | 49686 | 95 | 183 | 3083 | 11008 | 3657 | 6437 | 7749 | 10726 | | 2459 | |
| 1993 | 3487 | 6711 | 22602 | 87407 | 22 | 43 | 1742 | 6112 | 3218 | 5658 | 5260 | 5980 | | 2231 | |
| 1994 | 6600 | 12908 | 18098 | 32992 | 169 | 310 | 2573 | 9303 | 1743 | 3071 | 3659 | 4155 | | 1346 | |
| 1995 | 4171 | 8199 | 30324 | 44094 | 384 | 576 | 1946 | 7081 | 2532 | 4460 | 3728 | 4289 | | 1748 | |
| 1996 | 6026 | 11929 | 16317 | 28035 | 394 | 591 | 4217 | 15526 | 3571 | 6283 | 5535 | 6365 | | 2407 | |
| 1997 | 3828 | 7535 | 14711 | 24521 | 387 | 581 | 4443 | 16384 | 1550 | 2726 | 3210 | 3678 | | 1611 | |
| 1998 | 2595 | 5015 | 15207 | 28060 | 385 | 577 | 2669 | 9839 | 1359 | 1867 | 2032 | 2437 | | 1526 | |
| 1999 | 2738 | 5269 | 14585 | 23026 | 383 | 575 | 2022 | 7581 | 1709 | 2350 | 2734 | 3090 | | 1168 | |
| 2000 | 3493 | 6785 | 15950 | 24778 | 378 | 566 | 1905 | 7244 | 1315 | 1809 | 1189 | 1430 | | 533 | |
| 2001 | 5815 | 11449 | 22082 | 29875 | 376 | 564 | 2194 | 8335 | 1980 | 2724 | 2113 | 2501 | | 797 | |
| 2002 | 3592 | 6985 | 11094 | 18077 | 372 | 557 | 1595 | 6100 | 749 | 1029 | 639 | 752 | | 526 | |
| 2003 | 6072 | 11966 | 18783 | 28749 | 371 | 557 | 3091 | 11732 | 1952 | 2682 | 1128 | 1289 | | 1199 | |
| 2004 | 4623 | 9055 | 18589 | 32435 | 367 | 550 | 3427 | 13100 | 1302 | 1789 | 1402 | 1698 | | 1316 | |
| 2005 | 5265 | 10348 | 17008 | 30057 | 373 | 560 | 2879 | 10682 | 860 | 1177 | 890 | 1121 | | 994 | |
| 2006 | 3924 | 7651 | 18805 | 32890 | 392 | 587 | 2746 | 10356 | 1559 | 2141 | 997 | 1276 | | 1030 | |
| 2007 | 6565 | 12957 | 16018 | 25734 | 412 | 618 | 1757 | 7037 | 701 | 959 | 689 | 841 | | 958 | |
| 2008 | 4382 | 8572 | 10377 | 19761 | 429 | 644 | 2623 | 10755 | 1928 | 2650 | 858 | 1105 | | 1799 | |
| 2009 | 6074 | 11970 | 17065 | 27543 | 402 | 602 | 1979 | 8335 | 1034 | 1418 | 1678 | 2158 | | 2095 | |
| 2010 | 4581 | 8972 | 15301 | 22708 | 439 | 658 | 2662 | 11005 | 1061 | 1451 | 1117 | 1398 | | 1098 | |
| 2011 | 11177 | 22223 | 24960 | 57144 | 653 | 980 | 4718 | 19229 | 1504 | 2065 | 2598 | 3421 | | 3087 | |
| 2012 | 4969 | 9750 | 11411 | 21661 | 653 | 980 | 1080 | 4733 | 788 | 1075 | 335 | 422 | | 913 | |
| 2013 | 8190 | 16222 | 10860 | 22992 | 719 | 1077 | 2430 | 10112 | 2184 | 3002 | 503 | 660 | | 525 | |
| 2014 | 4165 | 8134 | 8320 | 16883 | 491 | 735 | 1542 | 6892 | 417 | 571 | 230 | 299 | | 340 | |

Appendix 4.x. (continued). Input data for small salmon returns to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

| Year of return to rivers | Returns of small salmon | | | | | | | | | | | | | |
|--------------------------|-------------------------|------------|------------|------------|------------|------------|------------|------------|---------------|---------------|------------|------------|----------------|------|
| | SFA 15 | | SFA 16 | | SFA 17 | | SFA 18 | | SFA 19-21 | | SFA 23 | | USA | |
| | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Point estimate | |
| Winbugs labels | SF15Sm_L[] | SF15Sm_H[] | SF16Sm_L[] | SF16Sm_H[] | SF17Sm_L[] | SF17Sm_H[] | SF18Sm_L[] | SF18Sm_H[] | SF19_21Sm_L[] | SF19_21Sm_H[] | SF23Sm_L[] | SF23Sm_H[] | USASm[] | |
| 1970 | 2834 | 6279 | 47779 | 67697 | 0 | 0 | 264 | 1073 | 16177 | 24106 | 5306 | 7521 | 0 | 0 |
| 1971 | 2113 | 4681 | 38388 | 54120 | 0 | 0 | 65 | 265 | 11911 | 18004 | 3248 | 4541 | 32 | 32 |
| 1972 | 2185 | 4699 | 48886 | 69270 | 0 | 0 | 131 | 530 | 11587 | 17992 | 1831 | 2506 | 18 | 18 |
| 1973 | 3010 | 6668 | 47190 | 66835 | 5 | 9 | 516 | 2095 | 14169 | 22159 | 5474 | 7012 | 23 | 23 |
| 1974 | 2226 | 4895 | 78091 | 110470 | 0 | 0 | 187 | 757 | 25032 | 39058 | 10195 | 12901 | 55 | 55 |
| 1975 | 2393 | 5298 | 69993 | 98443 | 0 | 0 | 112 | 454 | 10860 | 15753 | 18022 | 23101 | 84 | 84 |
| 1976 | 8667 | 14696 | 96504 | 136107 | 14 | 28 | 299 | 1212 | 21071 | 33009 | 22835 | 28864 | 186 | 186 |
| 1977 | 6085 | 12084 | 30621 | 42689 | 0 | 0 | 215 | 871 | 24599 | 37314 | 13738 | 16671 | 75 | 75 |
| 1978 | 4350 | 7749 | 29783 | 39927 | 0 | 0 | 78 | 316 | 7621 | 10023 | 6271 | 7695 | 155 | 155 |
| 1979 | 4378 | 9495 | 50667 | 70714 | 2 | 5 | 1857 | 7536 | 24298 | 37514 | 15356 | 20517 | 250 | 250 |
| 1980 | 7994 | 15278 | 41687 | 58839 | 12 | 23 | 520 | 2108 | 34377 | 50250 | 25139 | 31483 | 818 | 818 |
| 1981 | 9380 | 17119 | 63278 | 108226 | 259 | 498 | 2797 | 11348 | 21204 | 48945 | 16826 | 21803 | 1130 | 1130 |
| 1982 | 6541 | 13383 | 78072 | 133171 | 175 | 336 | 2150 | 8722 | 17619 | 27075 | 11811 | 15636 | 334 | 334 |
| 1983 | 2723 | 4638 | 24585 | 41332 | 17 | 32 | 212 | 858 | 9313 | 14068 | 9270 | 12592 | 295 | 295 |
| 1984 | 12003 | 15867 | 28714 | 49595 | 17 | 32 | 460 | 1667 | 18382 | 29867 | 15556 | 21678 | 598 | 598 |
| 1985 | 7003 | 15516 | 53393 | 92224 | 113 | 217 | 730 | 3167 | 24384 | 39541 | 13056 | 17928 | 392 | 392 |
| 1986 | 10813 | 23926 | 103230 | 178295 | 566 | 1088 | 965 | 3854 | 24369 | 39663 | 14274 | 20183 | 758 | 758 |
| 1987 | 9630 | 21220 | 74485 | 128644 | 1141 | 2194 | 1646 | 5713 | 27269 | 44266 | 13358 | 17662 | 1128 | 1128 |
| 1988 | 13168 | 29092 | 107071 | 184904 | 1542 | 2963 | 1381 | 4833 | 24509 | 39750 | 16381 | 23084 | 992 | 992 |
| 1989 | 6357 | 13900 | 66069 | 114097 | 400 | 770 | 893 | 3208 | 25602 | 41557 | 17579 | 24521 | 1258 | 1258 |
| 1990 | 7880 | 17314 | 73020 | 126115 | 1842 | 3536 | 983 | 3528 | 29471 | 48039 | 13820 | 19176 | 687 | 687 |
| 1991 | 4441 | 9828 | 53453 | 92327 | 1576 | 3028 | 1160 | 4166 | 9762 | 15955 | 13041 | 17685 | 310 | 310 |
| 1992 | 8853 | 19614 | 142416 | 204708 | 1873 | 3999 | 994 | 3531 | 13754 | 22269 | 13563 | 18404 | 1194 | 1194 |
| 1993 | 5783 | 12812 | 70090 | 175096 | 1277 | 2454 | 1146 | 3892 | 13297 | 21681 | 7610 | 8828 | 466 | 466 |
| 1994 | 9136 | 20208 | 41773 | 59888 | 210 | 385 | 671 | 2425 | 3154 | 5393 | 5770 | 6610 | 436 | 436 |
| 1995 | 2902 | 6429 | 44357 | 63453 | 658 | 987 | 543 | 1985 | 8397 | 13873 | 8265 | 9458 | 213 | 213 |
| 1996 | 6034 | 13370 | 32067 | 45995 | 710 | 1065 | 2431 | 8958 | 13120 | 22293 | 12907 | 15256 | 651 | 651 |
| 1997 | 5797 | 12845 | 14377 | 24122 | 517 | 776 | 561 | 2134 | 3410 | 5863 | 4508 | 4979 | 365 | 365 |
| 1998 | 6288 | 13932 | 21965 | 32523 | 508 | 762 | 633 | 2419 | 8833 | 11927 | 9203 | 10801 | 403 | 403 |
| 1999 | 4936 | 10929 | 21494 | 29707 | 413 | 620 | 705 | 2681 | 3971 | 5337 | 5508 | 6366 | 419 | 419 |
| 2000 | 7459 | 16520 | 31923 | 42435 | 395 | 593 | 615 | 2428 | 6155 | 8312 | 4796 | 5453 | 270 | 270 |
| 2001 | 4947 | 10953 | 26496 | 36655 | 415 | 622 | 822 | 3205 | 2326 | 3138 | 2513 | 2862 | 266 | 266 |
| 2002 | 11719 | 25958 | 40432 | 54790 | 390 | 585 | 844 | 3319 | 5197 | 7015 | 3501 | 3991 | 450 | 450 |
| 2003 | 3119 | 6904 | 26530 | 39772 | 515 | 773 | 773 | 3088 | 2844 | 3837 | 2292 | 2716 | 237 | 237 |
| 2004 | 12091 | 26783 | 43242 | 62082 | 330 | 495 | 1092 | 4339 | 3847 | 5192 | 3454 | 4297 | 319 | 319 |
| 2005 | 4117 | 9116 | 28441 | 47190 | 343 | 514 | 781 | 3015 | 2870 | 3871 | 3597 | 4640 | 319 | 319 |
| 2006 | 8724 | 19322 | 30671 | 52560 | 331 | 497 | 869 | 3406 | 5144 | 6940 | 3720 | 4743 | 450 | 450 |
| 2007 | 4259 | 9430 | 23038 | 44016 | 275 | 413 | 718 | 2820 | 4198 | 5664 | 2466 | 3136 | 297 | 297 |
| 2008 | 13601 | 30129 | 25722 | 46587 | 298 | 447 | 1245 | 5061 | 7282 | 9831 | 5924 | 7691 | 814 | 814 |
| 2009 | 5169 | 11445 | 10819 | 21456 | 233 | 350 | 302 | 1417 | 2066 | 2788 | 1603 | 2027 | 241 | 241 |
| 2010 | 8187 | 18132 | 48123 | 67747 | 258 | 387 | 877 | 3672 | 3686 | 4975 | 9114 | 11994 | 525 | 525 |
| 2011 | 10234 | 22668 | 39511 | 67884 | 291 | 436 | 1248 | 5124 | 3615 | 4878 | 4466 | 5943 | 1080 | 1080 |
| 2012 | 4350 | 9631 | 6914 | 13254 | 291 | 436 | 211 | 1077 | 346 | 466 | 178 | 219 | 26 | 26 |
| 2013 | 4661 | 10920 | 10106 | 20216 | 274 | 410 | 429 | 3006 | 922 | 1244 | 894 | 1151 | 78 | 78 |
| 2014 | 3786 | 8380 | 6457 | 11181 | 222 | 332 | 233 | 1987 | 529 | 713 | 677 | 910 | 110 | 110 |

Appendix 4.x. (continued). Input data for 2SW salmon spawners to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

| | Spawners of 2SW | | | | | | | | | | | | | | |
|-----------------------------|-----------------|----------------|----------------|----------------|----------------|----------------|----------------|----------------|-------------------|-------------------|----------------|----------------|-------------------|--|--|
| Year of return to rivers | SFA 15 | | SFA 16 | | SFA 17 | | SFA 18 | | SFA 19-21 | | SFA 23 | | USA | | |
| | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Point estimate | | |
| Winbugs labels | SF15S2_ L[] | SF15S2_ H[] | SF16S2_ L[] | SF16S2_ H[] | SF17S2_ L[] | SF17S2_ H[] | SF18S2_ L[] | SF18S2_ H[] | SF19_21 S2_L[] | SF19_21 S2_H[] | SF23S2_ L[] | SF23S2_ H[] | USAS2[] | | |
| 1970 | 1156 | 3252 | 5346 | 8242 | 18 | 47 | 304 | 1587 | 2388 | 4234 | 1536 | 4846 | 0 | | |
| 1971 | 510 | 1434 | 6724 | 11354 | 0 | 0 | 133 | 694 | 1418 | 2513 | 3612 | 6576 | 490 | | |
| 1972 | 2367 | 6656 | 17031 | 31450 | 0 | 0 | 148 | 775 | 1616 | 2865 | 6472 | 9806 | 1038 | | |
| 1973 | 2873 | 8081 | 19277 | 33170 | 0 | 0 | 165 | 863 | 2246 | 3984 | 2752 | 4412 | 1100 | | |
| 1974 | 3620 | 10183 | 31192 | 52012 | 0 | 0 | 151 | 790 | 2878 | 5103 | 8123 | 12046 | 1147 | | |
| 1975 | 1769 | 4975 | 18536 | 31972 | 0 | 0 | 91 | 473 | 1987 | 3523 | 10987 | 16209 | 1942 | | |
| 1976 | 3530 | 9928 | 11842 | 22152 | 1 | 4 | 116 | 604 | 1935 | 3432 | 10071 | 15583 | 1126 | | |
| 1977 | 4412 | 12408 | 30623 | 54071 | 0 | 0 | 198 | 1033 | 2559 | 4539 | 12013 | 18568 | 643 | | |
| 1978 | 2622 | 7375 | 6998 | 13535 | 0 | 0 | 223 | 1166 | 1948 | 3455 | 5346 | 8076 | 3314 | | |
| 1979 | 527 | 1482 | 3000 | 5806 | 3 | 7 | 115 | 598 | 1419 | 2517 | 3772 | 5650 | 1509 | | |
| 1980 | 3440 | 9677 | 17667 | 30961 | 1 | 4 | 198 | 1033 | 4170 | 7394 | 12023 | 19005 | 4263 | | |
| 1981 | 1380 | 3880 | 2392 | 10515 | 36 | 73 | 196 | 1027 | 3631 | 6439 | 3642 | 7014 | 4334 | | |
| 1982 | 991 | 2786 | 8418 | 28619 | 8 | 23 | 253 | 1322 | 1158 | 2053 | 4475 | 7939 | 4643 | | |
| 1983 | 906 | 2547 | 5516 | 17586 | 15 | 30 | 210 | 1100 | 1579 | 2800 | 468 | 3561 | 1769 | | |
| 1984 | 2656 | 5402 | 11650 | 30889 | 13 | 26 | 259 | 1148 | 1416 | 2512 | 12280 | 18798 | 2547 | | |
| 1985 | 4514 | 9180 | 14019 | 37030 | 8 | 15 | 871 | 4369 | 6761 | 11990 | 11885 | 18624 | 4884 | | |
| 1986 | 7279 | 14804 | 20606 | 54630 | 5 | 11 | 2164 | 11213 | 6624 | 11748 | 7224 | 11280 | 5570 | | |
| 1987 | 4122 | 8383 | 11414 | 31114 | 66 | 128 | 2534 | 10189 | 3676 | 6519 | 5628 | 8597 | 2781 | | |
| 1988 | 6582 | 13386 | 13801 | 36355 | 96 | 185 | 2451 | 9954 | 4322 | 7664 | 3420 | 5248 | 3038 | | |
| 1989 | 3944 | 8021 | 8466 | 22739 | 149 | 287 | 2042 | 8397 | 4735 | 8396 | 6310 | 9158 | 2800 | | |
| 1990 | 3886 | 7903 | 13669 | 36039 | 284 | 545 | 1829 | 7491 | 3530 | 6260 | 4926 | 7292 | 4356 | | |
| 1991 | 2193 | 4460 | 14200 | 37251 | 188 | 361 | 2275 | 9399 | 2912 | 5165 | 6080 | 9158 | 2416 | | |
| 1992 | 3639 | 7400 | 20770 | 30116 | 95 | 183 | 2291 | 9324 | 2588 | 4589 | 5826 | 8633 | 2292 | | |
| 1993 | 1239 | 2521 | 15239 | 59907 | 22 | 43 | 1296 | 5180 | 2493 | 4421 | 3291 | 3654 | 2065 | | |
| 1994 | 3639 | 7401 | 13418 | 24653 | 166 | 307 | 1920 | 7907 | 1339 | 2375 | 2387 | 2680 | 1344 | | |
| 1995 | 2519 | 5124 | 25326 | 36949 | 380 | 576 | 1453 | 6022 | 2218 | 3934 | 3126 | 3652 | 1748 | | |
| 1996 | 3688 | 7502 | 10743 | 18662 | 388 | 591 | 3166 | 13262 | 2946 | 5224 | 4009 | 4585 | 2407 | | |
| 1997 | 2316 | 4710 | 8106 | 13754 | 385 | 581 | 3334 | 13988 | 1140 | 2022 | 2219 | 2565 | 1611 | | |
| 1998 | 1512 | 3076 | 6098 | 10548 | 382 | 577 | 2000 | 8390 | 915 | 1261 | 1068 | 1302 | 1526 | | |
| 1999 | 1581 | 3217 | 6589 | 10660 | 379 | 575 | 1523 | 6493 | 1409 | 1941 | 1934 | 2181 | 1168 | | |
| 2000 | 2057 | 4184 | 7262 | 11408 | 376 | 566 | 1438 | 6214 | 1072 | 1477 | 805 | 1004 | 1587 | | |
| 2001 | 3521 | 7161 | 13688 | 18674 | 374 | 564 | 1654 | 7143 | 1812 | 2497 | 1699 | 2008 | 1491 | | |
| 2002 | 2120 | 4312 | 5332 | 8808 | 371 | 557 | 1203 | 5230 | 378 | 521 | 317 | 356 | 511 | | |
| 2003 | 3683 | 7491 | 10593 | 16372 | 368 | 557 | 2333 | 10063 | 1834 | 2528 | 878 | 998 | 1192 | | |
| 2004 | 2770 | 5633 | 10144 | 17965 | 365 | 550 | 2581 | 11219 | 1017 | 1401 | 1238 | 1492 | 1283 | | |
| 2005 | 3175 | 6457 | 10755 | 19354 | 371 | 560 | 2162 | 9124 | 646 | 890 | 726 | 914 | 1088 | | |
| 2006 | 2329 | 4737 | 9336 | 16594 | 390 | 587 | 2062 | 8851 | 1248 | 1720 | 796 | 1023 | 1419 | | |
| 2007 | 3994 | 8124 | 8963 | 14644 | 409 | 618 | 1320 | 6023 | 587 | 809 | 530 | 633 | 1189 | | |
| 2008 | 2618 | 5325 | 5376 | 10584 | 429 | 644 | 1961 | 9180 | 1778 | 2450 | 736 | 953 | 2809 | | |
| 2009 | 3684 | 7494 | 10062 | 16500 | 401 | 602 | 1481 | 7122 | 811 | 1118 | 1391 | 1774 | 2292 | | |
| 2010 | 2743 | 5580 | 7335 | 11078 | 438 | 658 | 1998 | 9419 | 910 | 1253 | 726 | 877 | 1482 | | |
| 2011 | 6902 | 14038 | 20445 | 47555 | 652 | 980 | 3543 | 16455 | 1467 | 2023 | 2430 | 3196 | 3872 | | |
| 2012 | 2988 | 6077 | 7603 | 14713 | 652 | 980 | 816 | 4070 | 601 | 828 | 238 | 298 | 2020 | | |
| 2013 | 5019 | 10208 | 6909 | 15204 | 989 | 1483 | 1826 | 8662 | 2057 | 2834 | 405 | 526 | 525 | | |
| 2014 | 2481 | 5045 | 5520 | 11240 | 709 | 1063 | 1161 | 5915 | 410 | 564 | 163 | 208 | 566 | | |

Appendix 4.x. (continued). Input data for large salmon spawners to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

| Year of return to rivers | Spawners of large salmon | | | | | | | | | | | | | | Point estimate |
|-----------------------------|--------------------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|----------------|----------------|-------------|-------------|----------|--|-------------------|
| | SFA 15 | | SFA 16 | | SFA 17 | | SFA 18 | | SFA 19-21 | | SFA 23 | | USA | | |
| | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | | | |
| Winbugs labels | SF15SLg_L[] | SF15SLg_H[] | SF16SLg_L[] | SF16SLg_H[] | SF17SLg_L[] | SF17SLg_H[] | SF18SLg_L[] | SF18SLg_H[] | SF19_21SLg_L[] | SF19_21SLg_H[] | SF23SLg_L[] | SF23SLg_H[] | USASLg[] | | |
| 1970 | 1779 | 5003 | 5790 | 8926 | 18 | 47 | 395 | 1824 | 3101 | 5499 | 1451 | 5705 | 0 | | |
| 1971 | 785 | 2207 | 7324 | 12369 | 0 | 0 | 173 | 797 | 1841 | 3264 | 3888 | 7405 | 490 | | |
| 1972 | 4011 | 11282 | 17648 | 32589 | 0 | 0 | 193 | 891 | 2099 | 3721 | 7246 | 10892 | 1038 | | |
| 1973 | 3883 | 10920 | 20126 | 34632 | 0 | 0 | 215 | 992 | 2612 | 4632 | 3050 | 4928 | 1100 | | |
| 1974 | 4960 | 13949 | 34352 | 57282 | 0 | 0 | 196 | 908 | 2878 | 5103 | 9090 | 13347 | 1147 | | |
| 1975 | 2239 | 6297 | 21355 | 36834 | 0 | 0 | 118 | 544 | 1987 | 3523 | 12335 | 17857 | 1942 | | |
| 1976 | 4644 | 13063 | 13867 | 25940 | 1 | 4 | 151 | 694 | 1995 | 3538 | 11183 | 17230 | 1126 | | |
| 1977 | 5315 | 14949 | 32337 | 57097 | 0 | 0 | 257 | 1137 | 3324 | 5895 | 13452 | 20470 | 643 | | |
| 1978 | 3496 | 9833 | 8128 | 15720 | 0 | 0 | 290 | 1340 | 2466 | 4373 | 5948 | 8955 | 3314 | | |
| 1979 | 1033 | 2906 | 4355 | 8426 | 3 | 7 | 149 | 688 | 1944 | 3448 | 4217 | 6264 | 1509 | | |
| 1980 | 4248 | 11947 | 18597 | 32590 | 1 | 4 | 257 | 1187 | 4849 | 8598 | 13190 | 21206 | 4263 | | |
| 1981 | 2935 | 8256 | 3586 | 15765 | 36 | 73 | 255 | 1131 | 4907 | 8702 | 3794 | 8056 | 4334 | | |
| 1982 | 1679 | 4723 | 10405 | 35376 | 8 | 23 | 329 | 1519 | 2464 | 4369 | 4903 | 9059 | 4643 | | |
| 1983 | 1535 | 4317 | 6852 | 21846 | 15 | 30 | 273 | 1264 | 2506 | 4445 | 92 | 4419 | 1769 | | |
| 1984 | 3362 | 6838 | 12341 | 32721 | 13 | 26 | 337 | 1320 | 1940 | 3441 | 13675 | 20961 | 2547 | | |
| 1985 | 7164 | 14571 | 16114 | 42563 | 8 | 15 | 1131 | 5010 | 8347 | 14803 | 13104 | 20811 | 4884 | | |
| 1986 | 9577 | 19479 | 24157 | 64044 | 5 | 11 | 2811 | 12889 | 10515 | 18647 | 8004 | 12623 | 5570 | | |
| 1987 | 6441 | 13099 | 14340 | 39088 | 66 | 128 | 3291 | 11711 | 5835 | 10347 | 6343 | 9594 | 2781 | | |
| 1988 | 9141 | 18592 | 16913 | 44553 | 96 | 185 | 3183 | 11442 | 7203 | 12773 | 3835 | 5787 | 3038 | | |
| 1989 | 6919 | 14072 | 12965 | 34822 | 149 | 287 | 2652 | 9651 | 6862 | 12168 | 7099 | 10086 | 2800 | | |
| 1990 | 5715 | 11623 | 22190 | 58504 | 284 | 545 | 2376 | 8611 | 6087 | 10793 | 5576 | 8051 | 4356 | | |
| 1991 | 4386 | 8920 | 23472 | 61572 | 188 | 361 | 2955 | 10803 | 4045 | 7173 | 6833 | 10180 | 2416 | | |
| 1992 | 6738 | 13704 | 33583 | 48697 | 95 | 183 | 2976 | 10717 | 3594 | 6374 | 6511 | 9488 | 2292 | | |
| 1993 | 3099 | 6302 | 22109 | 86914 | 22 | 43 | 1683 | 5953 | 3156 | 5596 | 4026 | 4746 | 2065 | | |
| 1994 | 6065 | 12334 | 17787 | 32682 | 166 | 307 | 2493 | 9088 | 1717 | 3045 | 2827 | 3273 | 1344 | | |
| 1995 | 3873 | 7877 | 30007 | 43778 | 380 | 676 | 1887 | 6922 | 2492 | 4420 | 3362 | 3923 | 1748 | | |
| 1996 | 5674 | 11541 | 15755 | 27367 | 388 | 691 | 4112 | 15244 | 3507 | 6219 | 4688 | 5497 | 2407 | | |
| 1997 | 3563 | 7247 | 13955 | 23677 | 385 | 581 | 4330 | 16078 | 1520 | 2696 | 2565 | 3028 | 1611 | | |
| 1998 | 2326 | 4732 | 14737 | 25493 | 382 | 577 | 2597 | 9643 | 1346 | 1854 | 1675 | 2074 | 1526 | | |
| 1999 | 2433 | 4948 | 13539 | 21905 | 379 | 575 | 1979 | 7464 | 1697 | 2338 | 2251 | 2601 | 1168 | | |
| 2000 | 3165 | 6437 | 15321 | 24069 | 376 | 566 | 1867 | 7142 | 1307 | 1801 | 975 | 1216 | 1587 | | |
| 2001 | 5417 | 11018 | 21201 | 28923 | 374 | 564 | 2148 | 8210 | 1970 | 2714 | 1831 | 2210 | 1491 | | |
| 2002 | 3261 | 6633 | 10618 | 17538 | 371 | 557 | 1562 | 6011 | 741 | 1021 | 442 | 542 | 511 | | |
| 2003 | 5666 | 11525 | 18102 | 27978 | 368 | 557 | 3029 | 11567 | 1931 | 2661 | 919 | 1074 | 1192 | | |
| 2004 | 4261 | 8666 | 17796 | 31517 | 365 | 550 | 3351 | 12895 | 1287 | 1774 | 1287 | 1574 | 1283 | | |
| 2005 | 4884 | 9934 | 16172 | 29104 | 371 | 560 | 2807 | 10487 | 839 | 1156 | 791 | 1012 | 1088 | | |
| 2006 | 3583 | 7288 | 17954 | 31911 | 390 | 587 | 2678 | 10174 | 1541 | 2123 | 847 | 1113 | 1419 | | |
| 2007 | 6145 | 12488 | 15191 | 24820 | 409 | 618 | 1715 | 6923 | 683 | 941 | 586 | 726 | 1189 | | |
| 2008 | 4028 | 8192 | 9601 | 18901 | 429 | 644 | 2547 | 10551 | 1912 | 2634 | 767 | 1007 | 2231 | | |
| 2009 | 5668 | 11529 | 16229 | 26612 | 401 | 602 | 1924 | 8186 | 1014 | 1398 | 1565 | 2034 | 2318 | | |
| 2010 | 4221 | 8584 | 14382 | 21722 | 438 | 658 | 2595 | 10826 | 1034 | 1424 | 996 | 1275 | 1502 | | |
| 2011 | 10619 | 21597 | 24053 | 55948 | 652 | 980 | 4601 | 18913 | 1482 | 2043 | 2532 | 3353 | 3914 | | |
| 2012 | 3230 | 6338 | 10861 | 21019 | 652 | 980 | 1059 | 4679 | 761 | 1048 | 300 | 387 | 2054 | | |
| 2013 | 7721 | 15704 | 10013 | 22035 | 717 | 1075 | 2372 | 9956 | 2165 | 2983 | 486 | 643 | 525 | | |
| 2014 | 3816 | 7762 | 8190 | 16676 | 489 | 733 | 1508 | 6799 | 410 | 564 | 220 | 289 | 572 | | |

Appendix 4.x. (continued). Input data for small salmon spawners to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

| Year of return to rivers | Spawners of small salmon | | | | | | | | | | | | USA Point estimate |
|-----------------------------|--------------------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|----------------|----------------|-------------|-------------|--------------------------|
| | SFA 15 | | SFA 16 | | SFA 17 | | SFA 18 | | SFA 19-21 | | SFA 23 | | |
| | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | |
| Winbugs labels | SF15SSm_L[] | SF15SSm_H[] | SF16SSm_L[] | SF16SSm_H[] | SF17SSm_L[] | SF17SSm_H[] | SF18SSm_L[] | SF18SSm_H[] | SF19_21SSm_L[] | SF19_21SSm_H[] | SF23SSm_L[] | SF23SSm_H[] | USASSm[] |
| 1970 | 1417 | 4396 | 25958 | 45876 | 0 | 0 | 167 | 842 | 9429 | 17358 | 3886 | 6101 | 0 |
| 1971 | 1056 | 3277 | 22463 | 38195 | 0 | 0 | 41 | 208 | 7246 | 13339 | 1216 | 2509 | 29 |
| 1972 | 1034 | 3208 | 27639 | 48023 | 0 | 0 | 82 | 416 | 7616 | 14021 | 0 | 1 | 17 |
| 1973 | 1505 | 4668 | 31703 | 51349 | 3 | 7 | 325 | 1645 | 9502 | 17492 | 4037 | 5575 | 13 |
| 1974 | 1098 | 3405 | 57376 | 89755 | 0 | 0 | 118 | 595 | 16680 | 30706 | 8071 | 10777 | 40 |
| 1975 | 1195 | 3707 | 50438 | 78888 | 0 | 0 | 71 | 357 | 5819 | 10712 | 15363 | 20442 | 67 |
| 1976 | 2480 | 7692 | 64526 | 104130 | 8 | 22 | 188 | 951 | 14196 | 26134 | 17572 | 23601 | 151 |
| 1977 | 2467 | 7653 | 13270 | 25338 | 0 | 0 | 135 | 634 | 15120 | 27835 | 9196 | 12129 | 54 |
| 1978 | 1398 | 4337 | 14689 | 24833 | 0 | 0 | 49 | 248 | 2857 | 5259 | 4256 | 5680 | 127 |
| 1979 | 2104 | 6528 | 31829 | 51876 | 1 | 4 | 1170 | 5915 | 15716 | 28932 | 11640 | 16801 | 247 |
| 1980 | 2996 | 9293 | 27791 | 44943 | 7 | 18 | 327 | 1655 | 18876 | 34749 | 19597 | 25941 | 722 |
| 1981 | 3183 | 9874 | 35423 | 80370 | 151 | 390 | 1762 | 8908 | 21096 | 38837 | 7805 | 12782 | 1009 |
| 1982 | 3038 | 9027 | 51324 | 106423 | 102 | 263 | 1354 | 6847 | 11244 | 20700 | 6532 | 10357 | 290 |
| 1983 | 820 | 2486 | 13298 | 30045 | 10 | 25 | 133 | 674 | 5653 | 10408 | 5132 | 8454 | 255 |
| 1984 | 1620 | 4971 | 7389 | 28271 | 10 | 25 | 177 | 1200 | 13658 | 25143 | 10290 | 16412 | 540 |
| 1985 | 3557 | 10936 | 32275 | 71106 | 66 | 170 | 145 | 1788 | 18024 | 33181 | 8164 | 13036 | 363 |
| 1986 | 5589 | 16990 | 71918 | 146983 | 330 | 852 | 63 | 1729 | 18187 | 33481 | 10725 | 16634 | 660 |
| 1987 | 4867 | 14920 | 49971 | 104131 | 665 | 1718 | 527 | 3075 | 20213 | 37210 | 10257 | 14561 | 1087 |
| 1988 | 6664 | 20468 | 71967 | 149800 | 899 | 2320 | 344 | 2388 | 18125 | 33366 | 13061 | 19764 | 923 |
| 1989 | 3191 | 9741 | 37696 | 85724 | 233 | 603 | 232 | 1650 | 18973 | 34928 | 13124 | 20066 | 1080 |
| 1990 | 3996 | 12190 | 46902 | 99996 | 1074 | 2771 | 229 | 1750 | 22080 | 40648 | 10025 | 15381 | 617 |
| 1991 | 2215 | 6872 | 39648 | 78522 | 919 | 2371 | 271 | 2068 | 7363 | 13556 | 9495 | 14139 | 235 |
| 1992 | 4426 | 13728 | 116657 | 178949 | 1092 | 2818 | 189 | 1634 | 10125 | 18640 | 9485 | 14326 | 1124 |
| 1993 | 2891 | 8968 | 52050 | 157056 | 745 | 1922 | 261 | 1805 | 9970 | 18354 | 5762 | 6868 | 444 |
| 1994 | 4554 | 14125 | 25649 | 43764 | 118 | 292 | 179 | 1266 | 2661 | 4900 | 4965 | 5738 | 427 |
| 1995 | 1451 | 4501 | 34650 | 53746 | 250 | 375 | 148 | 1055 | 6512 | 11988 | 8025 | 9218 | 213 |
| 1996 | 3017 | 9359 | 19511 | 29260 | 258 | 587 | 1005 | 5596 | 10909 | 20082 | 11576 | 13892 | 651 |
| 1997 | 2899 | 8991 | 8702 | 15524 | 256 | 384 | 203 | 1290 | 2917 | 5370 | 3971 | 4433 | 365 |
| 1998 | 3144 | 9752 | 13997 | 21887 | 255 | 382 | 228 | 1464 | 8818 | 11912 | 8775 | 10348 | 403 |
| 1999 | 2465 | 7646 | 12193 | 17943 | 253 | 380 | 347 | 1837 | 3895 | 5261 | 5196 | 6048 | 419 |
| 2000 | 3727 | 11560 | 18837 | 26196 | 252 | 378 | 314 | 1717 | 6148 | 8305 | 4455 | 5087 | 270 |
| 2001 | 2470 | 7663 | 15703 | 22815 | 250 | 376 | 403 | 2217 | 2315 | 3127 | 2210 | 2530 | 266 |
| 2002 | 5857 | 18166 | 25458 | 35509 | 249 | 373 | 426 | 2334 | 5180 | 6998 | 3232 | 3689 | 450 |
| 2003 | 1557 | 4829 | 15727 | 24997 | 248 | 371 | 396 | 2201 | 2829 | 3822 | 2069 | 2469 | 237 |
| 2004 | 6043 | 18744 | 27425 | 40613 | 246 | 369 | 496 | 2934 | 3833 | 5178 | 3229 | 4039 | 319 |
| 2005 | 2056 | 6377 | 17065 | 30189 | 246 | 368 | 300 | 1881 | 2854 | 3855 | 3433 | 4450 | 319 |
| 2006 | 4359 | 13522 | 19763 | 35085 | 247 | 370 | 358 | 2201 | 5119 | 6915 | 3528 | 4501 | 450 |
| 2007 | 2127 | 6587 | 14420 | 29105 | 248 | 372 | 330 | 1905 | 4176 | 5642 | 2305 | 2937 | 297 |
| 2008 | 6798 | 21086 | 16299 | 30904 | 249 | 373 | 451 | 3189 | 7252 | 9801 | 5729 | 7467 | 814 |
| 2009 | 2581 | 8007 | 5867 | 13313 | 233 | 350 | 105 | 953 | 2051 | 2773 | 1472 | 1864 | 241 |
| 2010 | 4090 | 12698 | 30506 | 44243 | 256 | 384 | 387 | 2516 | 3674 | 4963 | 9032 | 11901 | 525 |
| 2011 | 5114 | 15864 | 25264 | 45125 | 290 | 435 | 562 | 3506 | 3601 | 4864 | 4391 | 5867 | 1080 |
| 2012 | 2172 | 6738 | 3457 | 7895 | 290 | 435 | 119 | 860 | 343 | 463 | 167 | 208 | 26 |
| 2013 | 2328 | 7220 | 5360 | 12437 | 272 | 408 | 188 | 2095 | 919 | 1241 | 870 | 1127 | 78 |
| 2014 | 1890 | 5862 | 3893 | 7200 | 220 | 330 | 105 | 1505 | 527 | 711 | 669 | 900 | 110 |

Appendix 5: Model Walkthroughs

Summaries of the data preparation, model running and output processing were presented at a one-day workshop prior to the 2014 WGNAS meeting and provided step by step walkthroughs of the assessment processes. Where appropriate these have been updated in 2015.

NEAC pre-fishery abundance and national conservation limit model in R

[NB: Instructions apply to model version: "NEAC_PFA_CL_RR_model_2015-varM_v12" as used in 2015]

1) Introduction

This program performs the run-reconstruction estimation of pre-fishery abundance (PFA) of maturing and non-maturing 1SW salmon for each country (and region) in the NASCO-NEAC area. PFA is estimated for January 1st in the first sea winter. The program also establishes the pseudo stock-recruitment (S-R) relationship between lagged egg deposition and Total 1SW PFA, and applies a hockey-stick S-R analysis to estimate the National/Regional Conservation Limit where river-specific CLs are not available. The original model is described by Potter *et al.* (2004); minor changes to the estimation approach used for different countries and regions have been reported in the annual reports of WGNAS.

The model also estimates the proportion (mean and SE) of the Faroes catch originating in different countries/regions based on the genetic analysis of scales collected in the fishery between 1993 and 1995 and the estimated PFA for each country/region since 2001, when no fishery has been operating at Faroes. This requires the model to be run once to provide the catch proportions and then run a second time to provide a full PFA assessment. The catch proportions are also used in the Catch Options model.

2) To get started

- 2.1) Load RStudio or R;
- 2.2) Set up a folder from which you will run the program;
- 2.3) Use folder and file names without spaces;
- 2.4) Put the program, the input files (annual and multiannual) and the summary data file (see 6f) in this folder.

3) Input Data

- 3.1) Annual data (filenames: Annual-data-XX-YY.txt)
 - 3.1.1) There is a file for each country (XX) and region (YY) which contains the 40+ year time-series of data on catches, exploitation rates and non-reporting rates (plus additional data for some countries).
 - 3.1.2) To read the .txt files, it is easiest to open them from within Excel. i.e.
 - Open Excel;
 - select the correct folder;
 - click on 'Open'
 - You will probably need to change the setting in the lower right corner of the open box from 'Excel files' to 'All files';

- Double-click on the file you want to open and it should open the 'Text Import Wizard';
 - select 'finish' (If this doesn't work reopen the file, but select 'Delimited' at step 1, 'Tab' at step 2 and 'General' at step 3.)
- 3.1.3) Do not add any formatting to the file. If loading a new version of a file that has been saved in Excel (e.g. after addition of a new year's data), re-save the file by clicking 'Save As' and selecting 'Text (Tab delimited)' from the 'Save as type' list. This will remove the formatting and add the .txt extension.
- 3.1.4) Do not change the file name.
- 3.1.5) Close and save the file before running the programme. You will be prompted to confirm that you want to lose the formatting; click 'yes'.
- 3.2) Multiannual-data (file-name: 'Multiannual-data.txt' or similar)
- 3.2.1) This file contains most of the other parameters used in the model including: smolt age composition, fecundity and sex ratios by region, M, etc.
- 3.2.2) The second value listed is the 'lastdatayear' which needs to be updated to the latest year for which data are provided in the Annual-data-XX files.
- 3.2.3) The file is not formatted in columns so can be read easily in Notebook, which should be selected automatically if you click on the file to open it. (NB: If you open the file in Excel, don't save it because it will probably add " " marks to each line.
- 3.2.4) All blank lines and lines starting with '#' are ignored in this file. Apart from these:
- The first line must start with 'list(
 - The last line must be ')'
 - All other lines must be 'variable name' <- number, followed by a comma (except for the last data line which has no comma).
- 3.2.5) If the module estimating the composition of the Faroes catch is run (see below) the new values must be inserted at the end of the multiannual data file in place of the current ones.
- 3.2.6) Save the file before running the model.
- 4) Model structure
- 4.1) Introductory section: contains working directory, source files and various parameters controlling the way the program runs (some of these will need to be changed for your laptop (see below).
- 4.2) Functions: functions are sections of code that the program calls up to repeat the same job. They have to be run before they are first called by the program; this is achieved by placing them at the beginning of the code. The main functions run the hockey-stick analysis for the NCL model and output certain figures and tables.
- 4.3) Faroes and Greenland sections: these sections calculate the harvest in the distant water fisheries.

- 4.4) NEAC country/regions sections: there is a section for each country (in alphabetic order) and region to calculate the main outputs of the R-R model.
 - 4.5) Output summaries: this section creates NEAC summary figures and tables and the country/region data files for the Winbugs Forecast Model.
 - 4.6) Faroes catch composition: The final section estimates the proportion (mean and SE) of the Faroes catch originating in different countries/regions. This requires the model to be run once to provide the catch proportions and then run a second time to provide a full PFA assessment.
- 5) Running the code from RStudio
- 5.1) Open R Studio
 - 5.2) Select "File/Open File" and use the browser to select and open the code file; the code should open in the Top left panel. The code will have a name like "NEAC_PFA_CL_RR_model_2015_xxx"
 - 5.3) If you have been using the code recently, you can select "File/Recent Files" and select the file from the drop-down list (if it is there); you can open several code files simultaneously and they appear as tabs above the Top Left panel.
 - 5.4) To set up the code for your PC/laptop, R-click on the code and scroll down to:

line 40 -enter the full path name of the working directory (replace the text between the parentheses with the full pathname of the folder containing the code on your laptop (e.g. "D:/Modelling_NEAC/PFA_NCL_R/2014").

line 45 -ensure that the text between the parentheses shows the correct filename for the multi-annual data file.

lines 77-86 -select which countries you wish to run the assessment for by setting "run-XX": 1 = run country XX; 0 = do not run. The summaries will only be run if all countries are set to 1.

line 82 -set "PrintFigs" equal to '1' to output the summary figures (or any other value not to output them); otherwise "0".

line 89 -set "WinbugsFiles" equal to '1' to output the data files for the Bayesian forecast model (or any other value not to output them) ; otherwise "0"..

line 92 -set "PrintCountryTables" equal to '1' to output summary output data for each region that is run (or any other value not to output them) ; otherwise "0"..

line 98 -set "RunFaroeseCatchSplitEstimation" equal to "TRUE" to run the estimation of the Faroes catch composition; otherwise "FALSE".
 - 5.5) You do not need to save your changes before you run the code. [If you wish to save any changes, use "File/Save" or "File/Save As" as normal. It's a good idea to include the extension ".R". NB: You will be prompted to save the file before you close it.]
 - 5.5.1) To clear the 'console' area (lower left panel) press "Ctrl-L"

5.5.2) To run the program press “Ctrl-Alt-R”

5.5.3) You will see when part of the code run in console area. Errors will show in red. The run is complete when the final line shows “>”

6) Running the program from R

6.1) Open R Studio

6.2) Select “File/Open script” and use the browser to select and open the code file; the code should open in a separate panel. The code is currently called “NEAC_PFA_CL_RR_model_2014”

6.3) To set up the code for your PC/laptop, R-click on the code and scroll down to:

SET WORKING DIRECTORY (wd): In line starting “wd <-” replace the text between the parentheses with the full pathname of the folder containing the code on your laptop (e.g. “D:/Modelling_NEAC/PFA_NCL_R/2014”).

SET “run_XX”: in the lines starting “run_XX <-” select which countries you wish to run the assessment for by setting “run_XX”: 1 = run country XX; 0 = do not run. The summaries will only be run if all countries are set to 1.

SET ‘PrintFigs’: set “PrintFigs” equal to ‘1’ to output the summary figures (or any other value not to output them).

SET ‘WinbugsFiles’: set “WinbugsFiles” equal to ‘1’ to output the data files for the Bayesian forecast model (or any other value not to output them).

SET ‘PrintCountryTables’: set “PrintCountryTables” equal to ‘1’ to output summary output data for each region that is run (or any other value not to output them).

SET ‘RunFaroeseCatchSplitEstimation’: set “RunFaroeseCatchSplitEstimation” equal to “TRUE” to run the estimation of the Faroes catch composition; otherwise “FALSE” [SEE BELOW]

6.4) You do not need to save your changes before you run the code, but you may wish to save a version to be safe. To do this use “File/Save” or “File/Save As” as normal. It’s a good idea to include the extension “.R”. NB: You will be prompted to save the file before you close it.

6.5) To run the program select “Edit/run all”

6.6) You will see when the code runs in the ‘R console’ panel. Errors will show in red. The run is complete when the final line shows “>”

7) Running the Faroes stock composition

7.1) The ‘Multiannual-data’ file contains the latest estimates of the composition of the Faroes catch by European country/regions based on the results of the genetics analysis reported in the 2015 WGNAS report and the 2001-14 PFA outputs. These estimates may be updated if new genetics data are provided or additional years of PFA estimates are to be included.

- 7.2) To run the estimation, ensure that all “Annual-data-XX” files have been updated.
- 7.3) SET 'RunFaroeseCatchSplitEstimation' to “TRUE” and run the model.
- 7.4) The new stock composition parameters will be output in the file “Faroese_split_estimate.txt”; these data should be copied into the end of the “Multiannual-data” file to replace the values already there. These data are also required for the Catch Options models.
- 7.5) Reset SET 'RunFaroeseCatchSplitEstimation' to “FALSE” and run the model again to produce full updates of the PFA estimates.

8) Output files

The program produces the following outputs (if requested):

8.1) National plots: (filenames “Figure-XX”)

PDF files showing the national plots currently used in the WG report. This includes: maturing and non-maturing 1SW PFA; returns and spawners for 1SW and MSW; homewater exploitation rates; and total catches (incl. non-reported) for each country (XX). It also shows the pseudo stock–recruitment hockey-stock plots for each region; these show the estimated CL, where this is used in the assessment.

8.2) Regional data: (filenames “Region_data_XX_YY”)

Excel files showing PFA, returns, catch, exploitation rates and spawners for 1SW and MSW fish and total eggs and lagged egg estimates for each country (XX) and region (YY).

8.3) Input files for Forecast analysis: (filenames: “Winbugs_Data_XX_YY”)

Excel files for each country/region containing mean and sd estimates for the simulations for lagged eggs, 1SW returns and MSW returns.

8.4) Summary tables by country:

- Median spawner numbers
- Conservation limits and SERs
- Maturing 1SW PFA
- 1SW returns
- 1SW spawners
- Non-maturing 1SW PFA
- MSW returns
- MSW spawners

8.5) Summary plot for N-NEAC and S-NEAC

8.6) A formatted Excel workbook is set up to link to the output files and format the tables ready for use in the WGNAS report.

9) Common problems

9.1) The code will crash if an output file (Figure or Table) is left open. The error message (in red) may say:

- Error : cannot open file 'Fig-XX'

or

- Error in : cannot open file 'Region_data_XX.csv': Permission denied
- 9.2) It doesn't matter if an input file is open, but the program may not read the latest version if it has not been saved.
- 9.3) More problems to be added when they are found!

REPLACED