WGNAS Stock Annex for Atlantic salmon

Stock specific documentation of standard assessment procedures used by ICES.

Stock Atlantic salmon (sal-nea_SA)

Working Group on North Atlantic Salmon

(WGNAS)

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1 General

1.1 Stock definition

1.1.1 Background

Atlantic salmon, *Salmo salar* L., have a wide range of life-history strategies. Most forms are anadromous, however, with a juvenile phase in freshwater followed by a period at sea feeding and growing, during which the fish undergo extensive migrations in the open ocean, before they migrate back to freshwater to breed. Most Atlantic salmon return to their river of origin to spawn. This precise homing behaviour has resulted in groups of fish originating in different rivers or tributaries becoming genetically distinct as they adapt to the particular conditions that they face in their home river and along their migration routes. As a result, fish from one river or tributary can differ from fish originating in other rivers/ tributaries which have become adapted to a different set of conditions. These subgroups comprise genetically distinct 'populations' and these are regarded as the basic biological units of the Atlantic salmon species.

Large rivers and their tributaries can support several, genetically distinct populations, each with separate spawning areas within the main-stem of the river or its tributaries. In most instances, however, it is not possible to demarcate clear population boundaries within a river, and managing stocks and fisheries at this level of detail would be very complex. Thus, while there is a need to protect the sustainability of these units, the primary management unit (e.g. for reporting catch statistics and regulating fishing) is generally taken to be the river stock, comprising all fish originating in eggs laid within the river.

Atlantic salmon are native to the temperate and subarctic regions of the North Atlantic Ocean and there are over 2000 rivers draining into the North Atlantic that support the fish, about 1500 of which discharge into the Northeast Atlantic. In this area, salmon distribution extends from northern Portugal to northern Russia and Iceland, while in the Northwest Atlantic, the species ranges from northeastern USA (Connecticut) to northern Canada (Ungava Bay).

Ideally, the management of all individual river stocks, and the fisheries that exploit them, might be based upon the status of each individual population. This is not always practical, however, particularly where decisions relate to the management of distant water salmon fisheries, which exploit large numbers of stocks originating in broad geographic areas. WGNAS has therefore had to consider how populations or river stocks should be grouped in providing management advice. For this purpose, groups have been established which fall within the meaning of a stock as 'an exploited or managed unit' (Royce, 1984) and that are consistent with the ICES (1996)

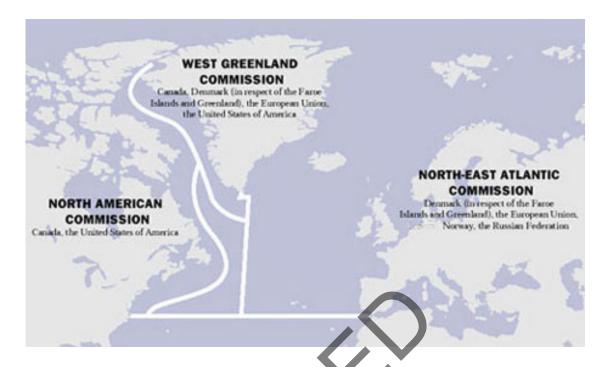
definition of salmon 'stocks' as 'units of a size (encompassing one or more populations) which provide a practical basis for the fishery manager'. The issues around the grouping of Atlantic salmon stocks for the provision of management advice are reviewed in detail in Crozier *et al.* (2003). Such stock groupings have typically been referred to as stock complexes.

Salmon mature at various sea ages, typically returning to freshwater to spawn after one to three years at sea, but also sometimes at older sea ages; this varies widely between populations. Those salmon that return after one year at sea are referred to as one-sea-winter (1SW) salmon, or grilse, with older fish categorised as 2SW, 3SW, etc. In practice, however, for management purposes these older sea age fish are typically aggregated and collectively referred to as multi-sea-winter (MSW) salmon. The sea age when salmon become sexually mature depends on genetics as well as growing conditions in the sea, and possibly freshwater, although the precise proximate factors initiating homeward migration are unknown (Hansen and Quinn, 1998). The sea age of Atlantic salmon is important in the context of stock definition since these different groups of fish have different migration routes, return at different times and are differentially exploited in fisheries. Thus, for example, it is only potential MSW salmon that are exploited in the distant water salmon fishery that operates off the west coast of Greenland.

1.1.2 Management framework for salmon in the North Atlantic

The advice generated by WGNAS is in response to Terms of Reference posed by the North Atlantic Salmon Conservation Organisation (NASCO), pursuant to its role in international management of salmon. NASCO was set up in 1984 by international convention (the Convention for the Conservation of Salmon in the North Atlantic Ocean), with a responsibility for the conservation, restoration, enhancement, and rational management of wild salmon in the North Atlantic. NASCO now has six Parties that are signatories to the Convention, Canada, Denmark (in respect of Faroe Islands and Greenland), the EU (which represents its Member States), Norway, Russia and the USA. While sovereign states retain their role in the regulation of salmon fisheries for salmon originating in their own rivers, fisheries within the jurisdiction of one Party that exploit salmon originating in the rivers of another Party may be regulated by NASCO under the terms of the Convention. This is currently the case for the distant water salmon fisheries at Greenland and Faroes.

NASCO discharges these responsibilities via three Commission areas shown below:



While homewater fisheries are not regulated directly by NASCO, national/ regional jurisdictions seek to comply with NASCO agreements and guidelines in exercising their responsibilities. In particular, NASCO's Agreement on the Adoption of a Precautionary Approach states that an objective for the management of salmon fisheries is to maintain the diversity and abundance of salmon stocks, and NASCO's Standing Committee on the Precautionary Approach interpreted this as being "to maintain both the productive capacity and diversity of salmon stocks" (NASCO, 1998).

NASCO's Action Plan for Application of the Precautionary Approach (NASCO, 1999) provides an interpretation of how this is to be achieved:

- "Management measures should be aimed at maintaining all stocks above their conservation limits by the use of management targets".
- "Socio-economic factors could be taken into account in applying the precautionary approach to fisheries management issues".
- "The precautionary approach is an integrated approach that requires, inter alia, that stock rebuilding programmes (including as appropriate, habitat improvements, stock enhancement, and fishery management actions) be developed for stocks that are below conservation limits".

In requesting scientific advice from ICES, NASCO asks for an annual review of events in the salmon fisheries and of the status of salmon stocks around the North Atlantic; NASCO also requests management advice for stocks in each of the Commission Areas. In fulfilling these requirements, three specific purposes have been identified for which stock groupings may be required (Crozier *et al.*, 2003):

- providing descriptions of the status of stocks;
- developing models to estimate and/or forecast pre-fishery abundance (PFA); and
- developing management advice for the distant water fisheries.

Crozier *et al.* (2003) further noted that there is no reason to assume that the same stock groupings should be used for all these purposes, indeed both the criteria used (e.g. geographical or biological features) and the resulting groups are likely to differ.

1.1.3 Stock groupings used by WGNAS in providing management advice

As noted above, Atlantic salmon would, ideally, be assessed and managed on the basis of river-specific stock units. In reality, <25% of the rivers with salmon populations in the North Atlantic are so assessed (Chaput, 2012; ICES, 2013). Consequently, stock status is often, of necessity, assessed at broader regional, national and subcontinental scales. While there might be merit in grouping stocks according to biological criteria (which could cross jurisdictional boundaries), it has generally been considered that the difficulties of collecting data in a similar format in different jurisdictions is likely to outweigh the benefits of using such groups (Crozier *et al.*, 2003). It is also recognised that compilations of data on stocks within each jurisdiction are of importance to regional / national managers. As such, regional / national stock groups are typically used by ICES in providing advice on the status of stocks, with additional information compiled on biological groups (e.g. sea ages) as required.

ICES has previously provided information on the status of stocks in the Northeast Atlantic Commission (NEAC) area by region or by country (as well as sea age). For the North American Commission (NAC) area similar information is provided for the USA and the five main provincial regions in eastern Canada: Labrador, Newfoundland, Québec, Gulf and Scotia-Fundy.

In providing management advice for the mixed-stock distant-water fisheries, broader scale stock groupings have been considered appropriate. For the NAC area this is based on the six geographic regions of North America detailed above. For the NEAC area, the following national groupings have been used in recent years to provide NASCO with catch advice or alternative management advice for the distant-water fisheries at West Greenland and Faroes.

SOUTHERN NEAC COUNTRIES	NORTHERN NEAC COUNTRIES
France	Russia
Ireland	Finland
UK (Northern Ireland)	Norway
UK (England & Wales)	Sweden
UK (Scotland)	Iceland (north/east regions) ¹
Iceland (south/west regions) ¹	

¹ The Iceland stock complex was split into two groups for stock assessment purposes in 2005 (ICES, 2005), largely on the basis of tag-recapture information. Prior to 2005, all regions of Iceland were considered part of the Northern NEAC stock complex.

These groups were deemed appropriate by WGNAS as they fulfilled an agreed set of criteria for defining stock groups for the provision of management advice that were considered in detail at the 2002 WGNAS meeting (ICES, 2002) and re-evaluated at the 2005 WGNAS meeting (ICES, 2005). ICES subsequently noted, however, that provision of catch advice for NEAC stocks in the distant water fisheries should preferably be based on a larger number of smaller management units, similar to those used in the NAC area (ICES, 2010a; 2011). Such an approach was developed at the

2013 WGNAS meeting (ICES, 2013) and indicative catch advice was provided at the country level as well as the Southern and Northern NEAC stock complexes. ICES is awaiting feedback from NASCO on the choice of management units.

Salmon from most NEAC stocks mix in the Norwegian Sea in autumn and winter, and are exploited by the fishery at Faroes. While there is evidence that some salmon from NAC rivers have been caught in the Norwegian Sea, they are currently not considered in the NEAC assessments, although this is now under review. Recent genetic information suggests that more North American fish than previously thought were exploited in the fishery at Faroes. Further details on the results of these investigations are provided in Section 3.3.3 of this report and potential options tor accounting for these fish in future catch advice is provided in Section 3.6. To date, consideration of the level of exploitation of national stocks in the Faroes fishery (ICES, 2005) resulted in the proposal that catch advice for the fishery should be based upon all NEAC area stocks and both 1SW and MSW fish.

In contrast, the fishery to the west of Greenland operates in an area where salmon from all North America and some Northeast Atlantic stocks mix in their second summer at sea. Catch advice for this fishery is thus based on non-maturing (potential MSW) fish from all regions of North America, while consideration of the level of exploitation of national stocks in the fishery from NEAC, resulted in catch advice being based upon only Southern NEAC non-maturing 1SW (potential MSW) fish (ICES, 2005).

1.2 Fisheries

Most exploitation of Atlantic salmon is restricted to fisheries close to or within the rivers of origin of the stocks; these homewater fisheries take adult fish that are mainly returning to these rivers to spawn. As noted above, these fisheries are not directly regulated by NASCO since the Parties retain responsibility for the regulation of fisheries for salmon originating in their own rivers. However, NASCO can regulate fisheries undertaken by a Party that take salmon originating in another Party's rivers, such as is the case for the distant-water fisheries at Greenland and Faroes. These fisheries take salmon originating in a large number of rivers over a wide geographical range.

1.2.1 The Northern Norwegian Sea Fishery

A longline fishery for salmon in parts of the Norwegian Sea, north of latitude 67°N, commenced in the early 1960s. Several countries participated in this fishery and the pattern of fishing, area of operation and catches changed markedly over the years. At its peak in 1970 this fishery harvested almost 1000 tonnes of salmon.

The Convention for the Conservation of Salmon in the North Atlantic Ocean, which resulted in the formation of NASCO, came into force in October 1983. The Convention created a large protected zone, free of targeted fisheries for Atlantic salmon in most areas beyond 12 nautical miles from the coast. An immediate effect was the cessation of the salmon fishery in the Northern Norwegian Sea outside the Faroes EEZ, with the last catches in this area reported in 1984 (ICES, 2013).

In the late 1980s and early 1990s, NASCO acted through diplomatic initiatives to address fishing for salmon in international waters by vessels registered to non-NASCO Parties. There have been no reports of such activities since.

1.2.2 The Faroes fishery

The fishery in the Faroes area commenced in 1968 with a small number of vessels fishing up to 70 miles north of the Faroes; initially catches increased slowly up to 40 tonnes in 1977. Danish vessels participated in the fishery between 1978 and 1982 and, at the same time, catches started to increase rapidly, peaking at 1025 tonnes in 1981. Several factors contributed to this increase: the season was extended, more vessels entered the fishery, and the fishery shifted northwards.

From 1982, the Faroese Government agreed to a voluntary quota system, involving a total catch of 750 tonnes in 1982 and 625 tonnes in 1983 (255 boats allowed 25 tonnes each). Since NASCO's establishment, regulatory measures or decisions have been agreed by the Northeast Atlantic Commission in most years (Table 1.2.2.1). These have resulted in greatly reduced allowable catches in the Faroese fishery, reflecting declining abundance of the salmon stocks. There has been no commercial salmon fishery targeting salmon around the Faroes since the early 1990s. Catches in the fishery are presented in Figure 1.2.2.1.

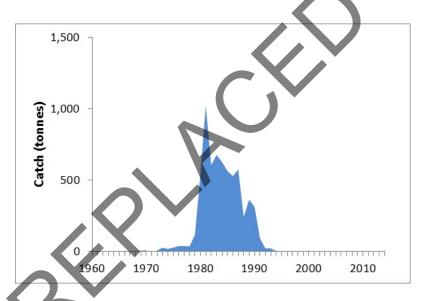


Figure 1.2.2.1. Nominal catch of salmon (tonnes, round fresh weight) in the Faroese longline fishery, 1960-2014.

Table 1.2.2.1. Summary of Regulatory Measures agreed by NASCO for the Faroese Salmon Fishery (courtesy of NASCO).

YEAR	Allowable CATCH (TONNES)	COMMENTS/OTHER DETAILS IN THE MEASURES/DECISIONS
1984– 1985	625	
1986	-	
1987– 1989	1790	Catch in any year not to exceed annual average (597t) by more than 5%.
1990– 1991	1100	Catch in any year not to exceed annual average (550t) by more than 15%.
1992	550	
1993	550	
1994	550	
1995	550	
1996	470	No more than 390 tonnes of the quota to be allocated if fishing licences issued.
1997	425	No more than 360 tonnes of the quota to be allocated if fishing licences issued.
1998	380	No more than 330 tonnes of the quota to be allocated if fishing licences issued.
1999	330	No more than 290 tonnes of the quota to be allocated if fishing licences issued.
2000	300	No more than 260 tonnes of the quota to be allocated if fishing licences issued.
2001– 2003	No quota set	It is the intention of the Faroese authorities to manage the fishery in a precautionary manner with a view to sustainability, and to make management decisions with due consideration to the advice from ICES concerning status of stocks contributing to the fishery.
2004– 2006	No quota set	It is the intention of the Faroese authorities to manage the fishery on the basis of the advice from ICES concerning status of stocks contributing to the fishery in a precautionary manner with a view to sustainability and taking into account relevant factors such as socioeconomic needs and other fisheries on mixed-stocks.
2007– 2012	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2013– 2015	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.

Note: The quotas for the Faroe Islands detailed above for the period 1984–2000 were agreed as part of effort limitation programmes (limiting the number of licences, season length and maximum number of boat fishing days) together with measures to minimise the capture of fish less than 60 cm in length. The measure for 1984/1985 did not set limits on the number of licences or the number of boat fishing days.

The Faroes salmon fishery operated over winter from November through to May. The salmon caught in the fishery originated almost entirely from European countries with salmon from many countries being present in the area (Jacobsen *et al.*, 2001). Small numbers of tagged fish originating in North America were also recaptured in the fishery (e.g. ICES, 1991), but excluded from catch advice. Genetic investigations, based on salmon scales removed from fish caught in the fishery in the 1980s and 1990s, suggested North American fish may make a larger contribution to the Faroes fishery than originally indicated (ICES, 2015). There was no consistent seasonal trend in the estimated proportion of North American fish in the catches at Faroes and so the overall percentages for 1SW (5.7%) and MSW (20.5%) salmon have been used in subsequent analyses. (ICES, 2015). WGNAS has been asked to consider the implications of the findings in providing future catch advice to NASCO.

The fishery exploited mainly 2SW fish, although some 1 and 3SW fish were also caught. Small salmon (<60 cm total length) in their first winter at sea were required to be discarded. Large numbers of farmed salmon were also observed at Faroes and there is evidence that farmed salmon escaping from net pens in Norway entered this area (Hansen *et al.*, 1987; Hansen and Jacobsen, 2003). Such farmed fish accounted for a significant proportion of the catch; in the early 1990s, the proportion of farmed fish in this area was estimated at between 25 and 40% (Hansen *et al.*, 1999).

Tagging studies (of adult fish caught in the fishery) have indicated that some fish caught at Faroes were apparently on their way westwards, as they were reported from West Greenland later the same year (Jákupsstovu, 1988). However, salmon tagged at West Greenland were also reported in the area north of the Faroes the following year (ICES, 1984). Thus, salmon of European origin are believed to move through the Faroese area on their way to the feeding areas in the West Atlantic as well as on their return to homewaters,

1.2.3 The Greenland fishery

Limited fishing at West Greenland is reported as far back as the early 1900s, although the present fishery dates from 1959 when local fishermen began setting fixed gillnets from small boats in certain fjords around Maniitsoq (Shearer, 1992). Rapid expansion along the coast followed and from the mid-1960s Faroese and Norwegian fishermen introduced offshore driftnets, followed soon by fishermen from Greenland and Denmark. At around the same time improvements in gear (the introduction of light monofilament nets) enabled fishing in daylight and improved the efficiency of the gear. As a consequence, catches rose quickly reaching a peak of almost 2700 tonnes in 1971. Fishing by non-Greenlandic vessels was phased out in 1972–1975. However, the total catch remained at around 2000 tonnes until 1976 when a TAC of 1190 tonnes was set; the fishery has been regulated since this time. Small catches of salmon are also made on the east coast of Greenland although these are sporadic and restricted by the small number of communities in this area and by drifting polar ice.

Regulatory measures have been agreed by the West Greenland Commission for most of the years since NASCO's establishment (Table 1.2.3.1). These have resulted in greatly reduced allowable catches in the West Greenland fishery, reflecting declining abundance of the contributing salmon stocks. In all but two years since 1998, the fishery has been restricted to an internal-use fishery and commercial export of salmon is not permitted. Catches in the Greenland fishery are presented in Figure 1.2.3.1.

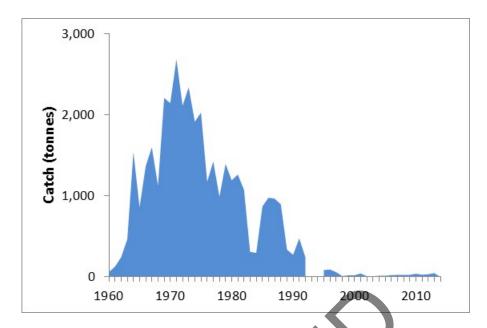


Figure 1.2.3.1. Nominal catch of salmon (tonnes, round fresh weight) in the Greenland salmon fishery, 1960–2013.

The Greenland salmon fishery operates in summer, with a fairly large proportion of the catch commonly being taken in the weeks after the opening of the season in August. Both drift and fixed gillnets continue to operate. The salmon caught in the fishery to the west of Greenland originate in both North America and the Northeast Atlantic. Data on continent of origin in the catch indicate a reasonably even split between fish from North America and Europe in the early 1990s (ICES, 2013). However, the proportion of North American fish in the catch has increased steadily since this time with North American fish comprising 80–90% of the fish caught in recent years.

The salmon caught at West Greenland are almost exclusively fish in their second summer at sea, however, these are non-maturing 1SW salmon destined to return to homewaters as 2SW, or older, fish. Fish from all parts of North America are taken in the fishery, while it is primarily only potential MSW salmon from southern countries in Europe (UK, Ireland and France) that are exploited here. Very few salmon of farmed origin appear in the catches at Greenland, and these are not taken into account in assessments.

Table 1.2.3.1. Summary of Regulatory Measures agreed by NASCO for the West Greenland Salmon Fishery (courtesy of NASCO).

YEAR	Allowable catch (tonnes)	COMMENTS/OTHER DETAILS IN THE MEASURES
1984	870	
1985	-	Greenlandic authorities unilaterally established quota of 852 t.
1986	850	Catch limit adjusted for season commencing after 1 August.
1987	850	Catch limit adjusted for season commencing after 1 August.
1988-1990	2520	Annual catch in any year not to exceed annual average (840 t) by more than 10%. Catch limit adjusted for season commencing after 1 August.
1991	-	Greenlandic authorities unilaterally established quota of 840 t.
1992	-	No TAC imposed by Greenlandic authorities but if the catch in first 14 days of the season had been higher compared to the previous year a TAC would have been imposed.
1993	213	An agreement detailing a mechanism for establishing annual quota in each of the years 1993 to 1997 was adopted by the Commission.
1994	159	
1995	77	
1996	-	Greenlandic authorities unilaterally established a quota of 174 t.
1997	57	An addendum to the 1993 Agreement was agreed by the Commission.
1998	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
1999	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
2000	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t. A Resolution Regarding the Fishing of Salmon at West Greenland was agreed by the Commission.
2001	28–200	Under an <i>ad hoc</i> management programme the allowable catch will be determined on the basis of cpue data obtained during the fishery.
2002	20–55	Under an <i>ad hoc</i> management programme the allowable catch will be determined on the basis of cpue data obtained during the fishery.
2003–2008	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
2009–2011	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
2012–2014	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be $20\ t.$
		Unilateral decision made by Greenland to allow factory landing with a 35 t quota for factory landings only in 2012 and 2013, 30 t in 2014, fishery restricted to catches used for internal consumption in Greenland, and higher catch figures based on sampling programme information are used for the assessments
2015	Internal consumption fishery only	Unilateral decision made by Greenland to set a 45 t quota for all sectors of the fishery, fishery restricted to catches used for internal consumption in Greenland, and higher catch figures based on sampling programme information and phone surveys are used for the assessments

1.3 Ecosystem aspects

Over the past 20 to 30 years, there has been a marked decline in the abundance of Atlantic salmon across the species' distributional range. Wild Atlantic salmon populations are declining across most of their home range and, in some cases, disappearing (ICES, 2008). Generally, populations on the southern edge of the distribution seem to have suffered the greatest decline (Parrish et al., 1998; Jonsson and Jonsson, 2009; Vøllestad et al., 2009). This may be linked to climatic factors. The decline in salmon abundance has coincided with a variety of environmental changes linked to an increase in greenhouse gases and a corresponding increase in temperatures (IPCC, 2001), which is most likely to have manifest effects at the edge of the species range. However, these areas are often also the ones with higher human population densities and therefore, typically, where potential impacts on the freshwater environment may also be greater. A range of factors in freshwater are known to affect stocks including, for example, contaminants, river obstructions, and changing river flows and temperatures (ICES, 2009b; 2010b; Russell et al., 2012). Such factors have potential implications for the survival of juvenile salmon and their resulting fitness when they migrate to sea as smolts (e.g. Fairchild et al., 2002).

Atlantic salmon occupy three aquatic habitats during their life cycle: freshwater, estuarine and marine. Similar factors contribute to mortality in each of these habitats competition, predation and environmental factors - but despite occurring in different habitats, these are not independent. Conditions experienced within the freshwater environment can affect the survival of emigrating smolts and marine conditions may subsequently modify the spawning success of fish in freshwater.

The decline in salmon populations has occurred despite significant reductions in exploitation, although this does not preclude possible fishery effects. An underlying cause has been a marked increase in the natural mortality of salmon at sea; the proportion of fish surviving between the smolts' seaward migration and their return to freshwater as adult fish (e.g. Peyronnet *et al.*, 2008; Chaput, 2012). For many stocks, return rates are now at the lowest levels in the time-series, even after the closure of marine fisheries. This reduced survival is thought to reflect climatic factors and broad-scale changes in ocean ecosystems as well as factors in freshwater. The exact processes controlling marine survival are relatively poorly understood (Friedland, 1998), although there is growing support for the hypotheses that survival and recruitment is mediated by growth during the post-smolt year, for European stocks at least (Friedland *et al.*, 2009).

Although their habitats are widely separated geographically, there is strong coherence in recruitment patterns between North American and European stock complexes. Recent research suggests recruitment is correlated with ocean temperature variation associated with the Atlantic Multidecadal Oscillation (AMO) (Friedland *et al.*, 2013). It further appears that there are differences in the mechanisms affecting stocks in the Northwest and Northeast Atlantic, with ocean climate variability during the first spring months of post-smolt life most important to the survival of North American stocks, while summer climate variation appears to be more important to adult recruitment variation for European stocks (Friedland *et al.*, 2013). It has been speculated that this may be related to the varying roles of predation pressure and size-related mortality on the two continental stock complexes.

In addition to changes in climate and potential issues operating in both freshwater and marine environments, various other factors have been postulated as possibly contributing to the decline in stock abundance, including predation, aquaculture

impacts and the effects of fisheries. Huge increases in aquaculture production of Atlantic salmon over recent decades (see Section 2.2.1 of the WGNAS report) have created some concerns for wild populations. The main potential impacts include: (i) genetic impacts on wild fish; (ii) discharge of organic material and other wastes; (iii) transmission of diseases and parasites (particularly sea lice) to wild populations; and (iv) concerns about obtaining adequate feed resources from an already heavily exploited marine ecosystem. For example, recent investigations in Norway have demonstrated that gene pools of wild salmon populations in a number of rivers have been gradually changed through introgression of genetic material from escaped farmed salmon (Glover et *al.*, 2012; Glover et *al.*, 2013). Sea lice also continue to be regarded as a serious problem for wild salmonids (Skilbrei *et al.*, 2013; Krkošek *et al.*, 2013) affecting their survival and perhaps also their life-history characteristics (Vollset *et al.*, 2014).

As well as declines in abundance, changes in salmon life histories are also widely reported throughout the geographic range of the species, affecting factors such as sea age composition, size at age, age-at-maturity, condition, sex ratio and growth rates (e.g. Nicieza and Braña, 1993; Hutchings and Jones, 1998; Niemelä *et al.*, 2006; Peyronnet *et al.*, 2007; Aprahamian *et al.*, 2008; Todd *et al.*, 2008). Changes are also manifest in freshwater stages, affecting factors such as the size and growth of parr and the age of smolting (e.g. Davidson and Hazelwood, 2005; Jutila *et al.*, 2006) and run timing (Kennedy and Crozier, 2010; Otero *et al.*, 2013).

2. Data

2.1 Introduction

Assessment of Atlantic salmon differs from the approaches commonly adopted for other species, for example in respect of the need for at sea surveys and collection of commercial catch per unit of effort (cpue) data. Instead, the assessment of salmon is based mainly on data collected on individual river stocks (e.g. catches and counts of returning fish), which are raised and aggregated to provide estimates of the number of fish returning to homewaters for different stock groupings. These estimates are used, in turn, to estimate abundance at earlier points in the life cycle of the fish and to inform the development of catch advice.

The provision of management advice for the mixed-stock fisheries at Faroes and West Greenland is based on assessments of the status of stocks at broad geographic scales. The North American Commission (NAC) area is divided into six management units, and the Northeast Atlantic (NEAC) Commission area is divided into 19 regions. Assessment of the status of the stocks in these areas is based on estimates of the total abundance - the pre-fishery abundance (PFA) - of different cohorts of salmon at a stage before the distant water fisheries operate. PFA is defined as the cohorts of salmon maturing as 1SW and MSW fish that are alive prior to all the marine fisheries for 1SW salmon (Rago *et al.*, 1993a). The catch advice for the NEAC area is then provided for the northern (N-NEAC) and southern (S-NEAC) stock complexes and for countries.

The models to estimate the PFA of salmon from different areas are typically based on the catch in numbers of one-sea-winter (1SW) and multi-sea-winter (MSW) salmon in each country or region, which are then raised to take account of estimates of non-reported catches and exploitation rates on the two age groups. In some cases, particularly in the NAC area, returns to homewater are estimated by alternative methods, such as counts at fishways and counting fences, or from mark and

recapture studies. The estimates of fish numbers returning to homewaters are then raised to take account of the natural mortality (M) between the date that the fish are deemed to recruit to the particular fishery of interest and the midpoint of the timing of the respective national fisheries. A value of 0.03 per month is assumed for M (Section 3.2.3). The date of recruitment of NAC stocks (and thus the PFA date) is taken as August 1st in the second summer at sea because these fish are first exploited in the distant water fishery at West Greenland. However, NEAC stocks recruit to the Faroes fishery during their first sea winter and so PFA is calculated at January 1st (i.e. eight months earlier) for these stocks.

2.2 Input data for assessments-NEAC area

PFA for NEAC stocks is estimated using the run-reconstruction approach described by Potter *et al.* (2004). The model estimates the PFA of both maturing and non-maturing 1SW salmon because both stock components may be caught in the Faroes fishery, and data for both the Faroes and West Greenland fisheries are incorporated into the model.

In order to run the NEAC PFA model, each country provides time-series (beginning in 1971) of catch in numbers, non-reporting rates and exploitation rates for 1SW and MSW salmon. Best estimates and a measure of the uncertainty or error are provided for the non-reporting and exploitation rate data in order to obtain a measure of the uncertainty in the PFA estimates, since these data are commonly derived from expert opinion. The latest data input variables used in running the NEAC assessment are listed at Appendix 3.

In some instances, the above information has been supplied in two or more regional blocks per country. In these instances, the model output is provided for the regions and is also combined to provide one set of output data for the country as a whole.

The input data for Finland consists solely of catches from the River Tana/Teno. These comprise both Finnish and Norwegian net and rod catches, as the river marks the border between these countries. The Norwegian catches from the river are not included in the input data for Norway.

Where possible, when the input data are themselves derived from other data sources, the raw data are included in the model. This allows the uncertainty in these analyses to be incorporated into the modelling approach. Thus, the catch and sample data used to estimate the catches of Scottish fish in the northeast English coastal fishery are incorporated into the assessments for both UK (England and Wales) and UK (Scotland). For Greenland, catch data are input in the form of harvests (reported and unreported) in weight, along with data from the West Greenland sampling programme.

Descriptions of how the model input data have been derived are presented below for different countries (updated from Crozier *et al.*, 2003; ICES, 2002). The methods used to derive the PFA input data for NEAC countries and options for improving the data are also discussed in Crozier *et al.* (2003).

2.2.1 Median dates of return to homewater fisheries

NEAC stocks recruit to the Faroes fishery during their first sea winter and so the date of recruitment (and thus the PFA date) is calculated at January 1st. In deriving PFA from the estimates of fish numbers returning to homewaters, it is necessary to take account of natural mortality between the date that the fish recruit to the particular

fishery of interest and the midpoint of the timing of the respective national fisheries. The median return date for 1SW and MSW fish for each country/region are provided in the table below. Thus there is about a six to nine month period between the PFA date and the median time of return to homewaters for maturing 1SW fish and 17 to 20 months for non-maturing fish.

Table 2.2.1.1. Midpoint of recruitment to homewater fisheries for NEAC countries/regions.

NEAC COUNTRY/ REGION	1SW	MSW
Northern NEAC		
Russia - Pechora River	8	8
Russia - Archangel / Karelia	7.5	8
Russia - Kola / White Sea	8.5	7.5
Russia - Kola / Barents Sea	7	6.5
Finland	6.5	6
Iceland - north & east	7	6
Norway	8	5
Sweden	8.5	6.5
Southern NEAC		
Iceland - south & west	7	6
UK (Scotland - east)	8	5
UK (Scotland - west)	8	7
UK (N. Ireland - Fo area)	7	5.5
UK (N. Ireland - FB area)	6.5	6
Ireland	8	5
UK (England & Wales)	8	5
France	8.5	8.5

2.2.2 Data inputs for Northern NEAC countries

2.2.2.1 Finland

Catch: The catch input to the model of Finland represents an estimate based on catch enquiries and the total number of licences issued. The Norwegian catch from the River Teno has been included in the Finnish catch, which results in a set of input data that effectively represents a single river system. Catch composition is estimated based on catch samples and corresponding scale analyses.

Level of unreported catch: Unreported catch is estimated by extrapolating the catches of the fishermen that failed to report their catches, as reporting is not mandatory.

Exploitation rates: Exploitation rates in the river fisheries are derived from radio tagging studies in 1992–1993 and 1995, when 70–100 adult fish (1SW and MSW) were tagged yearly in the estuary. Most of the important river fisheries were covered by these experiments.

2.2.2.2 Norway

Area split: Salmon catches in Norway are split into four regions on the basis of climatic differences and oceanographic differences among the areas. The areas are: (1) southeast Norway from the Swedish border to the border between Rogaland and Hordaland counties, (2) southwest Norway from the border between Rogaland and Hordaland counties to Stad (3) mid Norway from Stad to Lofoten, and (4) north Norway from Lofoten to the border with Russia.

Catch: Nominal catches of salmon in the four regions are used. In recent years there have been improvements in declaring catches. From 1979 there was a weight split 1SW/MSW (<3 kg/>3 kg). From 1993 the split was changed to 1SW/2SW/3SW (<3 kg/3–7 kg/>7 kg). Mean weight was provided for most groups and used to estimate numbers in the early part of the time-series, but in recent years the reported nominal catch (reported number of killed salmon in river and sea fisheries summed) is being used. In the input to the PFA model salmon smaller than 3 kg are regarded as 1 SW fish, whereas salmon larger than 3 kg are regarded as MSW fish. The two largest size groups are thus summed into MSW salmon. In the PFA model input the Norwegian catch data for the River Teno have been removed from the Norwegian catches and incorporated in the Finnish catches.

Unreported catch: No systematic effort is used to estimate unreported catches. Inputs are guesstimates based on occasional reports from test fishing, surveillance reports, and questionnaires. Currently there is no evidence that the level of unreported catches differs between the four regions. These estimates are provided by the management authorities.

Exploitation rates: The rates for the national model are guesstimates. For parts of southeast and southwest Norway they are derived from estimated marine exploitation rates from the River Imsa and the River Drammen. In recent years (from 2009 onwards) exploitation rates for many rivers (>50) have been taken into consideration. These exploitation rates have been obtained using a multitude of methods, mainly from drift counts of spawners or results from counting facilities combined with reported catches in the rivers. The exploitation rates have been adjusted in relation to reduced fishing effort. At present different exploitation levels are used for the different regions, reflecting different harvest regimes in the regions.

For Norway, only data from 1983 onwards have been used for assessment purposes.

2.2.2.3 Russia

Area split: The Atlantic salmon rivers of northwest Russia are split into the following four regions: Kola Peninsula - Barents Sea basin; Kola Peninsula - White Sea basin; Archangelsk Region and the Karelia; and the Pechora River region. The split is based on four regions with separate catch statistics and different biological characteristics of the stocks. For example, the difference in age composition and relative abundance of summer and autumn salmon evident among these four regions has influenced the split.

Catch: The declared catch data, in numbers, is available for the full time period (1971 onwards) for all four regions. Catches were allocated to 1SW or MSW age groups on the basis of commercial and scientific catch sampling programmes.

Level of unreported catch: Unreported catches in legal fisheries are estimated from logbooks and catch statistics, by comparing catch survey results with reported catch. Illegal catch is guesstimated and based on local knowledge of fisheries. The major

component of the illegal catch in the Barents Sea basin (Kola Peninsula and Pechora River) comes from in-river fisheries and a considerable part of the illegal catch in the White Sea basin (Kola Peninsula and Archangelsk region) comes from coastal areas and this contributes the greatest uncertainties. There is a particular problem with illegal catches on the Pechora River where scientific sampling programmes suggest that the illegal catch on this river is very high. The level of non-reporting increased considerably in the early 1990s due to the economic changes in Russia and temporary reduction of control and enforcement. Since late 2000s the higher level of non-reporting occurred in recreational fisheries due to unclear legislation for reporting. All these factors have been considered in deriving the level of unreported catch for the PFA model.

Exploitation rates: Information on exploitation rates is derived from several fisheries in the Kola Peninsula where counting fences are operated and from mark–recapture exercises on the rivers with recreational fisheries. Exploitation rates in Archangelsk and Pechora are guesstimated. These are the basis of the inputs to the model, regional sea age differences being adjusted on the basis of local knowledge from estimated stock levels.

2.2.2.4 Sweden

Catch: The catch input to the model is based on annual reported commercial salmon catch on the Swedish west coast, and on voluntary reporting from sport fishing in rivers. This reporting is detailed and considered accurate and is handled by the government agency "Swedish Agency for Marine and Water Management" (commercial catches) and the Swedish University of Agricultural Sciences (noncommercial catches). Unfortunately, reporting of catches from non-commercial fishing for salmon with gillnets or rod and line on the coast is lacking. However, due to fishing regulations these catches are small (permits required for trapnets, ban on gillnets in deeper waters, restrictions on the use of gillnets in shallow waters, limited fishing season, large marine protected areas, ban on selling fish, etc.).

There is a large proportion (ca. 60–70%) of reared fish in catches and stocks as a result of compensatory releases of reared smolts (ranching). As all ranched salmon are finclipped the catches of reared fish can be treated separately in the catch statistics. In the reporting from the commercial fishing the catch is not separated into wild and reared fish. The proportion of wild salmon is instead estimated from catch statistics in nearby rivers. Stocking of reared salmon is done in three rivers; all of these also have wild stocks in tributaries.

Catch-and-release is practised in most rivers (only rod-and-line fishing allowed in rivers) but the extent of C&R is not always known. Reported catch is landed fish, but in most rivers a proportion of the fish is released back alive and any subsequent mortality is not accounted for.

Level of unreported catch: Unreported catch, i.e. non-commercial catch of salmon in the coastal area with gillnets and rod and line, is estimated from guesstimates based on expert judgement from regional fishery officers and the Swedish University of Agricultural Sciences. These estimates are supported with catch inventories carried out in 1999 (Thörnqvist, unpublished), 2004 (Swedish Agency for Marine and Water Management), 2008 (Thörnqvist, unpubl.). Generally, the unreported catch is estimated to be 5–10% of the reported catch.

Exploitation rates: Few fish counters are present and tagging data exist mainly for reared stocks, where the fishing pressure is higher than for wild stocks. Input for the

PFA model is based on guesstimates. In the index River Ätran, data on size and composition of the spawning run and estimates of exploitation are developed at present. Since 2000, a fish ladder with an automatic counter has provided data on the spawning run in this river. Counter data in combination with results from small-scale tagging in this river are used to provide estimates of exploitation rates. One problem is that exploitation rates differ considerably between rivers. During the period 2000–2014 the average exploitation rates for the Swedish stock as a whole have been estimated at 34% for 1SW and 39% for MSW. The exploitation rate has increased in 2011-2014 due to increased gillnet fishing on the coast. This fishery has been closed and exploitation is expected to decrease in future.

2.2.2.5 Iceland

Area split: The input data for the PFA model is divided into two areas. Rivers in the west and south of Iceland are combined into one area and rivers in the north and east into another. This is done on the basis of historic tag recoveries in ocean fisheries (which occurred in different areas) and different climate and oceanic conditions affecting the salmon life cycle, e.g. run-timing, smolt age, and sea age. The southern and western parts of Iceland fall within the NEAC southern area, while the northern and eastern parts of Iceland fall within the NEAC northern area.

Catch: Age-class information is available from individual catch records from logbooks used in the rod fishery. The division into sea age classes is based on a bimodal weight distribution. The 1SW females are <3.5 kg and 2SW females >3.5 kg, while 1SW males are <4 kg and 2SW >4 kg. Scale analyses have shown that the presence of salmon having spent more than two winters at sea and of previous spawners is uncommon and that the categorisation into 1SW and 2SW age classes by weight is regarded as fairly accurate. The net catches are recorded on a daily basis. The age split in the net fishery is derived from the weight distribution in the rod fishery from the same river system or from rivers in the same area.

In the River Ranga in southern Iceland substantial smolt releases have occurred since the early 1990s and have now reached a level of 300 000 to 500 000 smolts annually. Originally, the River Ranga had a small salmon stock with an annual catch of 10 to 90 fish until 1990. The river has very limited habitat for salmon production, but these 'ranched' fish now support a substantial rod fishery. The catch in the River Ranga comprised 23% (18–27%) of the total reported salmon rod catch in Iceland between 2009 and 2013. Since these fish are expected to have very low spawning success in the river they are excluded from the PFA catch input data.

Level of unreported catch: The fishing rights in Icelandic salmon rivers belong to landowners who must, by law, form a fishery association to manage the fishing right. The rod fishing rights are leased to the highest bidder. No ocean or estuary fisheries are allowed. The unreported catch was originally believed to be low with a guesstimate value of 2% applied. With increased use of midwater trawls in pelagic fisheries off the coast of Iceland, new information was provided which suggested an increased level of salmon bycatch. Based on a questionnaire survey, the value of unreported catch was therefore revised after 1995 to a value of 10% of the declared salmon catch. However, more recent analyses of DNA, as well as scale analyses, from salmon sampled as bycatch by Icelandic fishing vessels, indicates a low percentage of Icelandic salmon. Based on this, and other available information, a new estimate of unreported catch is now applied for Iceland at 4% of the declared catch for 1SW and MSW salmon.

Exploitation rates: Rates of rod exploitation are based on rivers with fish counters and catch records from logbooks. The estimates of exploitation are 40–50% for 1SW salmon and 50% to over 70% for 2SW salmon. The exploitation estimate for an inriver gillnet fishery is 39% to 52%, with a higher exploitation rate on larger fish. Information on the number of fish subject to catch and release in rod fisheries is also available from logbooks. The proportion of released fish has been increasing since 1996. The reduced exploitation due to catch and release is taken into account in the annual estimate of exploitation for both 1SW and 2SW stock components in the PFA model inputs.

Median return date of 1SW and MSW: Run timing can vary both between years and between areas. The median return date of 1SW and 2SW salmon in south and west Iceland is mid-June and early June respectively. The median date of return is later in the north and east of Iceland, mid-June for MSW and early July for 1SW salmon.

2.2.2.6 Denmark

The Working Group collects and routinely reports the annual catch of salmon taken in Denmark. However, the small Danish catches are not included in the assessment process used in developing catch advice for the distant water fisheries.

Catch: The catch input is based on continuously collected reports of salmon taken in the recreational fishery in Danish west coast streams (from Internet sources), which all hold populations of wild salmon. In four of these, where salmon populations have always been found, there is a large proportion of reared (finclipped) salmon in the catch, but these are all F1 offspring from the original populations. In the one catchment in eastern Denmark (Gudenå), where the salmon population is not genetically native to the stream, the annual catch is guesstimated.

Level of unreported catch: Unreported catch is expected to be negligible in the western streams because the fishing is closely regulated and controlled by the anglers. In the eastern stream (Gudenå) unreported catch is guesstimated.

Exploitation rates: Exploitation rates may be derived from the total catch related to estimates of the total run (calculated by mark–recapture surveys on a three-year cycle in the four streams with original populations on the west coast).

2.2.3 Data inputs for Southern NEAC countries

2.2.3.1 France

Catch: The estimation of salmon catch in France comes from two main sources: (1) mandatory declaration of rod and line catches and from the Adour nets operating in the lower river (scales are sampled from each fish caught) to the Office National de l'Eau et des Milieux Aquatiques (ONEMA), under the Ministère de l'Ecologie, which assumed responsibility from the Conseil Supérieur de la Pêche (CSP) in 2006; and (2) mandatory declaration of catches made by professional net fishermen to Affaires Maritimes, under the Ministère de la Mer, who since 2008 have delegated responsibility for collection and first processing of catch data to the Regional Boards for Sea Fisheries and Aquaculture Catch. At the same time, catches at sea are declared to the Institut Français de Recherches pour l'Exploitation de la Mer (Ifremer), who are responsible for archiving and scientific processing of all fisheries data. Salmon and sea trout catches have not been reliably collated and made available until recently. Since 1985, the 1SW/MSW split has been based on scale interpretation of the in-river catch (based on scale reading) and on a categorisation based on length thresholds for

catches in estuaries and at sea. The figures prior to 1985 are not considered as reliable as the later ones.

Level of unreported catch: Unreported legal catch for the rod and line fishery has been estimated by catch inquiries made by environmental inspectors of ONEMA on each river. These procedures are still operating in some areas, but estimates are considered less reliable in recent years. The estimation of the professional net fishery catch (Adour Basin) is thought to be reliable and no unreported legal catch is considered to apply.

For most years, the unreported illegal catch is not assessed and a minimal value is provided on a precautionary basis. This unreported illegal catch has been assessed in some years by *ad hoc* inquiries in the estuary of a number of rivers in Brittany (e.g. in 2001) and on the coast (e.g. Baie of Mont Saint-Michel in 2000). The "unreported catch" is included in the nominal catch. No estimates of unreported catch are available for the early part of the time-series (prior to 2001). Thus, the rates input to the model for 1SW and MSW for the early period are near zero and range from -0.00001 to 0.00001. Higher values in the range 20% to 40% for 1SW and 15% to 30% for MSW fish are applied more recently.

Exploitation rates: Exploitation rates are derived from the index River Scorff in Brittany. This is an in-river rate, by rods only, where there are no, or very few, fish thought to be caught on the estuary or coast. Rates are also derived for the Adour river system, where a rough estimation is provided by using the lower values of adult run estimates through facilities in the three rivers flowing to Adour, and the declared catches on the coast, estuary and river, respectively by nets and rods. Some caution is necessary regarding these rates from the Adour given the uncertainties in the different estimates. The rod catch on the index River Nivelle is very small and the probable net exploitation in the estuary and coast is unknown, so exploitation rates are not used for this system. Some data on exploitation rates are also collected by ONEMA on the index River Bresle, but sea trout are the dominant angled species in this river.

2.2.3.2 Ireland

Catch: The data are derived from annual declared catches within fisheries districts, management units implemented by Regional Fisheries Boards. Since 2007 river and estuarine specific angling and commercial catch data have been complied. The Fisheries Boards were amalgamated into a single body, Inland Fisheries Ireland, in 2010 which currently takes responsibility for compiling catch statistics. Catches are split by age on the basis of a reported age distribution from 1980 to 1988. In the absence of any other information the mean proportion of 2SW salmon in the series (7.5%) has been used since 1988 and a mean of 10% has been used prior to 1980. Since the introduction of a carcass tagging and logbook scheme for angling and commercial fisheries in Ireland in 2002, sea age classes in the time-series since 2007 have been determined based upon catch dates and weights in accordance with national river stock assessments. The catch is not corrected for returns from releases of smolts for ranching or enhancement but these are not a major component of the catch.

Level of unreported catch: The values are guesstimated from local reports and knowledge achieved during catch sampling and fisheries protection activities.

Exploitation rates: A coded-wire tagging (CWT) programme has been operated in several rivers in Ireland since 1980. Up to 300 000 hatchery smolts and up to 5000 wild smolts are tagged and released annually. There is also a substantial dataset on

wild salmon from the monitored River Burrishoole, providing a further index of wild returns and exploitation rates. Overall, there are estimates of exploitation rates available for three wild stocks and seven hatchery stocks for both 1SW and 2SW salmon. Up to the closure of the marine mixed-stock fishery in 2006, the annual mean of the 1SW wild exploitation index is used as the input data for the lower range of exploitation in the PFA model while the mean of the 1SW hatchery index is used as the upper range. The annual mean of the 2SW wild and hatchery exploitation index was used as the input data for the upper and lower range of exploitation in the PFA model depending on which is higher or lower in that year. Since 2006 the main exploitation input has been from the rod catch which is estimated from coded wire tags estimates for some rivers and also rivers with counters.

2.2.3.3 UK (England & Wales)

Catch: Nominal catches for UK (England & Wales) have been derived from the catch returns submitted by netsmen and anglers and split into 1SW and MSW categories using two different methods. Since 1992, monthly age—weight keys derived from salmon caught in the River Dee trap (an index river) have been used to estimate the age of all rod-caught fish where a weight and date of capture have been provided. This has then been scaled up to the total catch (rods and nets combined) on a pro-rata basis. In earlier years (1971–1991), the age composition of the total catch has been estimated using the mean weight of the fish caught and the mean weight of 1SW and MSW salmon recovered in tagging programmes.

As the contribution of farmed and ranched salmon to the national UK (England & Wales) catch is negligible, the occurrence of such fish is ignored in the assessments of the status of national stocks. However, a large proportion of the fish taken in the northeast coast fishery are destined for Scottish rivers, and these are deducted from the returning stock estimate for UK (England & Wales) and added to the data for UK (Scotland) in the ICES assessment. This proportion is estimated to have declined from 95% of the northeast net catch in the early part of the time-series to 75% in the late 1990s and to around 65% since 2003. This reflects both the steady improvement in the status of the stocks in northeast England and the phase out of the English driftnet fishery since 2003.

Level of unreported catch: All licence holders are required to provide the Environment Agency with details of their catch of salmon and migratory trout and the number of days fished on each river or, for nets, each fishery at the end of the season. Catch returns are received from all net licence holders and from ~90% of full season anglers, and the latter account for the majority of fish caught in a catchment, typically 96–98%. The main correction for underreporting is therefore currently made in respect of perceived inaccuracies in the returns, although more substantial corrections have applied in the past.

There are few independent measures of underreporting in the rod fishery, but these indicate that the level is currently small. A value of 10% is applied for correction purposes based on the method of Small (1991). Historically, underreporting was a much more serious problem. As a result of changes in the licensing and associated catch return system covering UK (England & Wales) in the early 1990s, the percentage of underreporting in the rod catch was estimated to have decreased from ~ 50% to ~20%. Since the mid-1990s, awareness campaigns and enhanced catch reminder systems have further reduced underreporting to the levels currently estimated.

For the net fishery, a figure of 8% has been used in recent years to adjust for the level of underreporting, based on the outcome of surveillance operations. The level may have been substantially higher in the past in certain fisheries, possibly as much as 50%. However, following the successful introduction of logbooks and a carcass tagging scheme in 2009, there is now considered to be minimal underreporting in net fisheries. A figure of 2% has been assumed since 2009.

An earlier questionnaire survey of Environment Agency enforcement staff suggested illegal catches were around 12% of the declared net and rod catch. However, since the introduction of a carcass tagging scheme and a ban on the sale of rod caught fish in 2009, it has been substantially more difficult to dispose of illegally caught fish. Since this time, illegal catches have been estimated to have been reduced to 6% of the declared catch.

Exploitation rates: Exploitation rates for a number of monitored fisheries in UK (England & Wales) are derived annually. National exploitation rates have then been estimated by deriving time-series of 'standard fishing units' employed in the salmon fisheries for the period 1971 to the present. For the period 1971 to 1997, these are calculated from the numbers of licences issued weighted by their relative catching power, which is estimated from historic cpue data; and for the period 1998 to the present, they are calculated from the numbers of days fished by different net categories weighted in the same way. The annual exploitation rates are then estimated by referencing the number of 'standard fishing units' employed over the two periods relative to average age-specific exploitation estimates derived for the 1997 and 1998 seasons.

Additional information: Further details on the derivation of estimates within UK (England & Wales) are available in the annual stock status reports (e.g. Cefas, Environment Agency and Natural Resources Wales, 2014), available at:

http://www.cefas.defra.gov.uk/publications/files/SalmonReport2013-final.pdf

2.2.3.4 UK (Northern Ireland)

Area split. Originally, a single assessment was carried out for UK (Northern Ireland). However, the data used were derived from two fishery management areas (Foyle and Fishery Conservancy Board (FCB) areas), which publish separate catch statistics and have differing fishing regulations. On the basis that stock status in the two areas may differ (Crozier *et al.*, 2003) the two areas were assessed separately from 2001.

Catch: As no commercial fishing has been conducted in the Loughs Agency area since 2010 and in the DCAL area since 2012 the Northern Ireland catch statistics currently (since 2014) rely solely on rod catches. Overall UK (Northern Ireland) rod catch estimates are available since the introduction of a carcass-tagging scheme in 2001. These catch statistics are used as an input in the model. Estimates of sea age composition of the catch for the time-series are based on 1SW/MSW data from adults returning to the River Bush (an index river).

Level of unreported catch: Estimates of unreported catch, as a result of illegal fishing, are based on intelligence reports from DCAL and Loughs Agency fishery officers. These are guesstimates only, with no verification possible. Annual adjustments in unreported catches have been used since tagging programmes started in the mid-1980s. Prior to that, a constant under reporting figure is used, as no annual data are available. The introduction of the carcass tagging scheme in 2001 has led to a reduction in unreported catches.

Exploitation rates: Estimates of exploitation rates were historically based on the River Bush microtagging programme. Exploitation from this monitored river (which is in the DCAL fishery area) was used as an input figure for all UK (Northern Ireland) fisheries (Foyle and DCAL areas). However, as currently no commercial fishery for salmon exists in the DCAL and Loughs Agency areas, exploitation rates are based on rod exploitation in the DCAL and Loughs Agency alone.

Possible improvements: A possible improvement would be to have better data available on sea age composition of all Northern Irish fish. Currently the River Bush estimate is applied to all Northern Irish data, but independent data from large river systems like the Bann and Foyle would result in a more reliable country wide estimate. In addition, a higher return rate for the carcass tagging scheme would result in more reliable estimates of exploitation rates. Recently the carcass tagging return rate in UK (Northern Ireland) has varied between 14% and 55%.

2.2.3.5 UK (Scotland)

Area split: The country is divided into eleven regions for the purposes of collating and publishing salmon and sea trout fishery statistics (Marine Scotland Science, 2012). Within the PFA run-reconstruction model, UK (Scotland) is divided into two broad areas (east and west), the split being influenced by the contrasts in climate, river size and the timing and sea age of returning fish. The east grouping comprises the east, northeast, Moray-Firth, and north statistical regions, the remaining regions comprise the west grouping.

Catch: Catches were collated according to the area split defined above. Reported retained catches of wild salmon, taken by both net and rod fisheries, are provided separately for two age classes, one sea-winter and multi sea-winter fish. Catch sampling programmes have shown that there is a variable (by region, year, and fishery) proportion of 1SW salmon categorised as MSW salmon in the reported catches. Catch data used in the model are not currently corrected to account for such misreporting.

Level of unreported catch: The ranges used in the national model are based on previous guesstimates made by local managers in some eastern areas of the country (MAFF, 1991). The differences in the ranges used for the east and west groupings are based on a subjective view that unreported catches in the west area are likely to be greater than in the east area. It is thought that human population densities are lower in the west and therefore there is likely to be less surveillance over the reporting of salmon catches. Further, west coast rivers are generally smaller and more numerous than east coast rivers, leading to a greater number of locations where unreported catches may be taken. Ranges provided are a subjective estimate of uncertainty in these parameters.

Exploitation rates: Rates for the national model are guesstimates derived from estimated exploitation rates over a range of fisheries in the early part of the timeseries (Shearer, 1992), and the time-series of exploitation rates derived from the River North Esk (Marine Scotland Science, 2010), together with reported net effort indices and the proportion of rod caught fish which were subsequently released (e.g. Marine Scotland Science, 2013).

2.2.3.6 Spain

The Working Group collects and routinely reports the annual catch of salmon taken in the recreational rod fisheries in Spain (mainly Asturias). However, the small

Spanish catches are not included in the assessment process used in developing catch advice for the distant water fisheries.

2.2.4 Data inputs for Faroes and West Greenland fisheries

2.2.4.1 Faroes

Reported catch: Catches are derived from the landings of salmon caught in the commercial and research fisheries that operated in the Faroes EEZ and the northern Norwegian Sea. Catches for each season (i.e. November in year n to May in year n+1) are assigned to the second year (i.e. year n+1). These fish are classified into 1SW and MSW age groups according to their age (or potential age) on January 1st during the fishery (i.e. a post-smolt caught in November is classified as 1SW).

Unreported catch: All fish less than 63 cm total length have been discarded in this fishery and so an unreporting rate of 10–15% (with an error of +/- 5%) has been used for 1SW fish; there is thought to have been negligible non-reporting of MSW fish.

Catch composition: Estimates of the proportion of farmed fish in the catch for the period 1981 to 1995 have been derived from scale reading (ICES, 1996; Hansen *et al.*, 1997); prior to 1981 all fish are assumed to have been wild, and since 1997 a value of 0.8 has been used.

Tagged fish originating in North America have been recaptured in the fishery (e.g. ICES, 1991), but excluded from catch advice.

The country of origin of the catch had been estimated based on tagging studies undertaken in the early 1990s (Hansen *et al.*, 1999). These were subsequently replaced by estimates based on genetic analysis.

Genetic investigations, based on salmon scales removed from fish caught in the fishery in the 1980s and 1990s provided an estimated proportion of North American fish in the catches at Faroes. Estimates of 5.7% (1SW) and 20.5% (MSW) have been used in subsequent analyses (ICES, 2015).

The composition of the European component was investigated using individual genetic assignments and gave an overall 1SW stock composition of 84.2% Southern European, 9.0% Northern Europe, 1.2% Icelandic and 5.7% North American (ICES, 2015). The overall composition of the MSW catch was determined as 20.9% Southern European, 58.0% Northern Europe, 0.6% Icelandic and 20.5% North American (ICES, 2015).

It was not possible to use the genetic assignments to estimate the composition of the catches to country/regional level, but they suggested that the composition within the stock complexes was broadly similar to the relative proportions of the PFA estimates and so the breakdown of catches at this level can be made by applying the relative proportions of PFA (ICES, 2015). Sources of uncertainty in these estimates are described in ICES (2015).

2.2.4.2 West Greenland

Catch: The total nominal catch (i.e. tonnes round fresh weight) in the West Greenland fishery is reported and converted to numbers using a mean weight obtained from the sampling programme.

Unreported catch: Estimates of unreported catch were not provided for the period from 1993 to 1999; an annual estimate of non-reported catch, varying from 5 to

20 tonnes was provided by the Greenland representative. Since 2000 a nominal figure of 10 t per year has been provided.

Efforts have been made to provide further information on the level of unreported catch at West Greenland. Since 2002, some assessment of the unreported catch, primarily for commercial landings, has been provided by comparing the weight of salmon seen by samplers involved in the international sampling programme and the corresponding community-specific reported landings. However, since sampling only occurs during a portion of the fishing season, these are considered to be minimum estimates for unreported catch. In addition, there is currently no quantitative approach for estimating the unreported catch for the private fishery. A telephone survey of fishers was carried out following the 2014 and 2015 seasons and provisional findings were provided to WGNAS in 2015 and 2016. These are discussed further in Section 5.1 of the Working Group report (see above); such investigations may provide a basis for revising estimates of unreported catch in future.

Continent of origin: The catch at West Greenland was divided into NAC and NEAC components using scale characteristics until around 2000 and since that time genetic analysis has been used. For the period when scale characteristics were used, the input data to the model are the minimum and maximum estimates of the proportion of NAC fish (from which minimum and maximum proportions of NEAC fish are calculated). For the subsequent period, the inputs are the numbers of NAC and NEAC fish identified in the samples.

2.2.5 Improvements to NEAC input data

NEAC countries have made ongoing efforts to improve the input data used in assessments. Modifications to input variables are reported by WGNAS in the year in which they are first implemented.

Over recent years, efforts have been made to reduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes). However, the methods used to derive estimates of unreported catch vary markedly between countries. For example, some countries include only illegally caught fish in the unreported catch, while other countries include estimates of unreported catch by legal gear as well as illegal catches in their estimates.

Descriptions of the national approaches used for evaluating unreported catches have been reported at various WGNAS meetings (e.g. ICES, 1996; 2000; 2002; 2010a). In addition, detailed reports describing national procedures for evaluating illegal and unreported catch, and efforts to minimise this, were submitted by parties to NASCO in 2007 in support of a special theme session on this issue. Full details are available at: http://www.nasco.int/pdf/2007%20papers/CNL(07)26.pdf.

Input data commonly rely on rod catches and the practice of catch and release has become increasingly important in recent years to reduce levels of exploitation on stocks. In the NEAC areas, catch and release estimates from the rod fisheries are not available from all countries and, when they are, corrections for catch and release mortality are commonly not applied. As the practice of catch and release is increasing, WGNAS have previously recommended (ICES, 2010a) that consideration should be given to incorporating mortality associated with this practice in river-specific, regional and national assessments.

The procedures currently used to incorporate catch and release and unreported catches into regional, national and international assessments are summarised in Appendix 1 (from ICES, 2010a).

One weakness of the NEAC model is that it is heavily dependent upon catch data and estimates of exploitation rate. In most salmon fisheries in the NEAC area, more than half the catch is reported, and in many cases it approaches 100%. However, as stocks have declined, exploitation rates have been reduced to very low levels, and estimates of abundance are therefore becoming increasingly sensitive to this parameter. This inevitably means that uncertainty in the estimates is increasing, and it therefore strengthens the need to make use of alternate sources of information on stock abundance, such as adult counts.

2.3 Input data for assessments-NAC area

The run-reconstruction model for NAC developed by Rago *et al.* (1993a) is used to estimate the PFA of non-maturing 1SW salmon of North American origin (beginning in 1971). Only the West Greenland fishery is of relevance in the context of distant water exploitation of NAC stocks. This fishery exploits predominantly (>95%) 1SW non-maturing salmon (destined to return primarily as 2SW salmon) and hence it is only necessary to estimate the abundance of this age group prior to the fishery at Greenland. The other fish taken in the fishery represent 2SW and older non-maturing salmon and previous spawners (ICES, 2003). However, PFA estimates for maturing 1SW salmon as well as large salmon (containing all MSW age groups of salmon including repeat spawners) can be derived from the run-reconstruction model.

The starting point for the reconstruction requires estimation of the returns of 2SW salmon and large salmon to the six regions in eastern North America: Labrador, Newfoundland, Québec, Gulf, Scotia-Fundy, and USA. With the progressive closure of commercial fisheries (1984 for the Gulf and Scotia-Fundy regions; 1992 for Newfoundland; and 1998–2000 for Labrador and Québec) abundance estimates have relied less on harvests and increasingly on estimated returns to rivers raised to production areas. The returns for each region are estimated with the uncertainty defined by a range of minimum and maximum values based on the best information available for each region (Chaput *et al.*, 2005).

The annual pre-fishery abundance of non-maturing 1SW fish for year i, destined to be 2SW returns (excluding 3SW and previous spawners), represents the estimated number of salmon at West Greenland prior to the start of the fishery on August 1st. Definitions of the input variables used in the model are given in Table 2.3.1. The PFA estimate is constructed by summing 2SW returns in year i+1 [NR2(i+1)], 2SW salmon catches in commercial and Aboriginal peoples' food fisheries in Canada [NC2(i+1)], and catches in year i from fisheries on non-maturing 1SW salmon in Canada [NC1(i)] and Greenland [NG1(i)].

Table 2.3.1. Definitions of key variables used in continental run-reconstruction models for North American salmon.

	T. I. C. DEA.
i	Index for PFA year corresponding to the year of the fishery on 1SW salmon in Greenland and Canada
M	Natural mortality rate (0.03 per month)
t1	Time between the midpoint of the Canadian fishery and return to river = 1 month
S1	Survival of 1SW salmon between the homewater fishery and return to river $\{exp^{\cdot M^*tt}\}$
H_s(i)	Number of "Small" salmon caught in Canada in year i; fish <2.7 kg
H_l(i)	Number of "Large" salmon caught in Canada in year i; fish >=2.7 kg
AH_s	Aboriginal and resident food harvests of small salmon in northern Labrador
AH_l	Aboriginal and resident food harvest of large salmon in northern Labrador
f_imm	Fraction of 1SW salmon that are immature, i.e. non-maturing: range = 0.1 to 0.2
af_imm	Fraction of 1SW salmon that are immature in native and resident food fisheries in N Lab
q	Fraction of 1SW salmon present in the large size market category; range = 0.1 to 0.3
MC1(i)	Harvest of maturing 1SW salmon in Newfoundland and Labrador in year i
i+1	Year of fishery on 2SW salmon in Canada
MR1(i)	Return estimates of maturing 1SW salmon in Atlantic Canada in year i
NN1(i)	Pre-fishery abundance (PFA) of non-maturing 1SW + maturing 2SW salmon in year 1
NR(i)	Return estimates of non-maturing + maturing 2SW salmon in year i
NR2(i+1)	Return estimates of maturing 2SW salmon in Canada
NC1(i)	Harvest of non-maturing 1SW salmon in Nfld + Labrador in year i
NC2(i+1)	Harvest of maturing 2SW salmon in Canada
NG(i)	Catch of 1SW North American origin salmon at Greenland
T2	Time between the start of the fishery at West Greenland (August 1) and return to the coast of North America = 10 months
S2	Survival of 2SW salmon between August 1 (at West Greenland) and return to the coast of North America $\{\exp^{\cdot M^*t_2}\}$
MN1(i)	Pre-fishery abundance of maturing 1SW salmon in year i

2.3.1 Data inputs for NAC

The latest data input variables used in running the NAC assessment are listed at Appendix 4. More detailed descriptions of how the model input data have been derived for each region of North America are presented below.

2.3.1.1 Labrador

For Labrador stocks, it was thought inappropriate to develop total recruits from angling catches and exploitation rates similar to techniques used for rivers in insular Newfoundland. The problem with using angling catches to derive returns for Labrador is, that until 1994, there were no estimates of exploitation rates available other than for the salmon population of Sand Hill River and these were 20 years out

of date. Also, because Labrador coastal rivers are isolated, the exploitation rates are low and highly variable depending on the presence of an angling camp and its success in attracting guests as well as the nearness of local communities. Thus, exploitation rates would depend, and vary from one year to the next, on the success of angling camps in attracting anglers and may not be applicable to other Labrador rivers. Thus, all estimates of returns and spawners until 1998 were based on commercial catches as the only source available of usable continuous time-series of data.

Before 1998

The general approach is to use exploitation rates to convert commercial catches of small and large salmon in Labrador to total population prior to the commercial fishery. River returns and spawners were estimated by subtracting the commercial catch from these populations, and accounting for non-Labrador interceptions. The estimated number of Labrador origin large returns is calculated as:

$$LR = (CC*PL) / u$$
 (1)

where,

LR = Labrador returns, PL = proportion Labrador origin, CC = commercial catch, and u = exploitation rate

The estimated number of Labrador origin small returns is determined from equation (1) but using commercial catches of small salmon.

Parameter values for sea age and the proportion of salmon of Labrador origin comes from the sampling program in the commercial fishery, 1974-1991. In 1997, commercial sampling resumed with samples being collected throughout the fishery at Makkovik and Rigolet in SFA 1 and Cartwright and St Lewis/Fox Harbour in SFA 2. River age distribution of commercial samples of small and large salmon from Labrador have been found to consist, on average, of about 75-80% river age 4 and older in SFAs 1 & 2. The commercial samples came from commercial catches sampled in Labrador at several sites along the Labrador coast including Square Islands (SFA 2) and at Nain (SFA 1) (Anon, 1993b). In total, 46 320 salmon were sampled for scales and aged. Labrador salmon stocks are thought to contribute about 70% of the total production of four year, and older, river age salmon, with the other 30% coming from northern Québec. Thus, when non-Labrador salmon are factored in at 30% applied to the river age distribution, then 60-80% of the harvest of small and large salmon (PL) in Labrador are of Labrador origin (Anon, 1993b). In 1997, in SFA 1, the percentage of the commercial catch that was of Labrador origin was for large salmon 68% (95% C. I. 64.3-72.5%); whereas for small salmon it was 39% (95% C.I. 35.6%-41.6%). In 1997, in SFA 2, the percentage of the commercial catch that was Labrador origin was for large salmon 92% (95% C. I. 88.4-95.2%); whereas for small salmon it was 80% (95% C.I. 74.8%–85.0%).

Exploitation rates (u) were calculated from the smolt tagging study in 1969–1973 on Sand Hill River (Reddin, 1981; Reddin and Dempson, 1989). Exploitation rates of 0.28 to 0.51 for small salmon and 0.83 to 0.97 for large salmon from the tagging study were changed to base exploitation rates of 0.3 to 0.5 on small salmon and 0.7 to 0.9 on large salmon and were assumed to apply to all of the salmon populations in SFAs 1, 2, and 14B for the period of 1969–1991 (Anon., 1993b). After 1991, due to the Management Plans for the commercial fishery in Labrador and Newfoundland, several changes occurred that would reduce exploitation of Labrador origin salmon. These changes

include: (1) reductions in effort as commercial salmon fishermen chose to sell their licences from a buy-out agreement begun in 1992, (2) a moratorium on commercial fishing in Newfoundland would increase the number of Labrador salmon in Labrador coastal waters, and (3) season reductions due to the varying opening dates and early closures from the quotas applied in 1995 and 1996. The effects of these changes were quantified in the exploitation model as follows:

$$u=1-e^{-aF}$$
 (2)

where: a = fraction of the 1991 licensed effort remaining in 1992–1996.

In 1994–1996, the licensed effort for all of Labrador was 37% of the 1991 level of 570 licences, in 1993 it was 55%, and in 1992 it was 87%. In any given year, it was assumed that 90% of licensed fishermen were active. Fishermen reported during public consultations that in 1995 and 1996 many licensed salmon fishermen did not fish for salmon in 1995-1996 but fished for crab instead. This was verified by Fisheries Officers who reported that of the 218 licensed salmon fishermen only 132 were active in 1996. Another method of obtaining actual effort information is also available since, beginning in 1993 commercial fishing vessel (CFV) numbers have been recorded on sales receipts issued to fishermen by fish plants. Enumeration of licensed salmon fishermen actively fishing was made by determining the number of CFVs in the Statistics Branch catch records. Active effort in 1991 and 1992 was assumed to be 90% as it was in 1993 and 1994 from the CFV file. Thus, the exploitation rates (u) were modified due to effort reductions in equation (2) using estimated active licences from 1991 as a base and the number of active licences in 1995, 1996 and 1997. The modified exploitation rates (ue) for 1992-1997 used the licensed effort in equation (2).

The tagging study on Sand Hill River, 1969–1973 showed that Labrador small and large salmon were not only caught in Labrador, but also in the commercial fisheries along the northeast coast of Newfoundland (both small and large) and at West Greenland (large only) (Anderson, 1985). For small salmon, out of a total of 100 (1SW) tag returns there were 24 from Newfoundland. For large salmon, out of a total of 137 (2SW) tag returns there were 41 from Newfoundland.

For 1992–1997: the moratorium on commercial fishing in Newfoundland would have released small and large salmon to Labrador. The effect of salmon released from Newfoundland in 1992–1996 was evaluated against the exploitation rates as follows:

un =
$$(1-((24 * (1 - ue))/100)) * ue$$
, for small salmon, and
un = $(1-((41 * (1 - ue))/137)) * ue$, for large salmon (3)

The new estimates of fishing mortality (un) in 1992–1994 included adjustments for the closure of the commercial fishery in Newfoundland based on the results of the Sand Hill River tagging study. Season reductions due to the varying opening dates and early closures from the quotas applied in 1995 and 1996. In 1995, adjustments were made to account for the new opening date for the commercial fishery in Labrador of July 3 changed from June 20 the previous year. For 1995, the accumulative effect of these, weighted to SFA catches, was to reduce the catch so that for small salmon the current catch represents 86.0% of small salmon and 62.7% of large salmon. In 1996, the opening date reverted to June 20 but the quota levels resulted in early closures in SFA 2 of 2A - July 10, 2B - July 8, and 2C - July 2 while SFA 1 and 14B did not close. For 1996, the accumulative effect of these weighted to SFA catches was to reduce the catch so that for small salmon the current catch represents 53% of small salmon and

61% of large salmon. In 1997, the opening date remained at June 20 but the quota levels resulted in early closures in SFA 2 of 2A -July 12, 2B - July 15, and 2C - July 13 while SFA 1 closed on October 15 as the quota was not caught. For 1997, the accumulative effect of these early closures was to reduce the catch so that for small salmon the current catch represents 47% of small salmon and 64% of large salmon. The season changes reduce catches and hence lower exploitation rates. The effect of shorter seasons in 1995, 1996 and 1997 was evaluated against the exploitation rates in section B as follows:

The new estimates of fishing mortality including effort reductions, adjustments for the closure of the commercial fishery in Newfoundland, and shorter seasons due to opening dates and quotas results in the following exploitation rates which were applied to catches. The cumulative effect of factors A, B, and C is to reduce exploitation on Labrador origin salmon.

Labrador origin 2SW returns (LR2SW) were derived from eq. 1 by

$$LR2SW = LR * P2SW$$
 (5)

where: P2SW = proportion of the large salmon that is 2SW salmon.

The SR1SW were calculated as in equation (5) but using P1SW which is the proportion of the catch that is 1-sea winter in age and maturing to enter freshwater and spawning in the year of capture. The parameter values for P1SW of 0.1 to 0.2 come from Anon. (1991).

The 2SW component was estimated separately for salmon caught in SFA 1, 2 and 14B. In SFA 1, commercial sampling at Nain of large salmon showed the proportion of 2SW were on average about 84% (n=6542), 1977–1991. Thus, a range of 0.7–0.9 was used for SFA 1. In SFA 2, commercial sampling of large salmon averaged 69% (n=4793) 2SW salmon, 1977–1991. There were no commercial samples available for SFA 14B. Thus, for SFAs 2 & 14B a range of 0.6–0.8 was used. For the 1SW component, commercial samples at Nain in SFA 1 of small salmon showed the proportion of 1SW salmon were on average about 94% (n=4757). In SFA 2 the 1SW component was on average about 97% (n=8872) of small salmon. There were no samples from commercial sampling in SFA 14B. In 1997, aged commercial samples indicated that the previous range was acceptable.

Total river returns of 2SW salmon (TRR) were calculated as follows:

$$TRR = LR2SW / (1-us)$$
 (6)

The total river returns of small salmon are also calculated by equation 6 but from SR. Spawning escapement (SE) or spawners was calculated according to the formula:

$$SE = TRR - AC, (7)$$

where:

AC = angling catch which includes retained catch plus 10% of hook & released mortality for released salmon.

A couple of modifications were made to the estimation procedure for Labrador in 1997. First, determination of exploitation rates were calculated separately for SFA 1, 2 and 14B using the active effort individually for each SFA. For SFA 1, the active number of licences declined from 141 in 1991 to 39 in 1997. For SFA 2, the active number of licences declined from 320 in 1991 to 99 in 1997. For SFA 14B, active licences declined from 52 in 1991 to 0 in 1997 when the fishery was closed. Exploitation rates determined as in equations 2, 3 and 4 are: SFA 1 - small was 0.0735 to 0.1399 and - large was 0.2221 to 0.3959; and SFA 2 - small was 0.0384 to 0.0728 and large was 0.1589 to 0.2799.

Numbers of small and large salmon for SFAs 1 & 2 were estimated from the exploitation model while for SFA 14B the results of assessments on Forteau Brook and Pinware River were expanded to include all the watersheds in SFA 14B. Returns to SFA 14B were 663 to 1545 small salmon and 146 to 327 large salmon.

Total mortalities of small and large salmon were accounted for by summing commercial catches of small salmon in Labrador and Newfoundland, large salmon in Labrador, Newfoundland, and Greenland, angling catches in Labrador of small and large salmon including 10% of the hook and released salmon, and small and large spawners. All of the above mortality estimates except catches of Labrador salmon in Newfoundland, 1969–1991 and Greenland could be obtained from equations 1 to 7. Catches in Newfoundland and Greenland were assessed as follows:

Greenland: for 1969–1992 and 1995–2004, removals of Labrador salmon by the Greenland fishery were assessed from data based on the sampling program in commercial fish plants at West Greenland (Anon, 1996). The Greenland fishery catches salmon that would have returned to homewaters as large salmon in the year following the Greenland fishery. Numbers of Labrador salmon were determined by converting catches in kg to numbers of salmon of 1SW North American origin that were of river age 4 and older. The number of Labrador salmon were estimated by assuming that 70% of the production of 4-year and older river age salmon are from Labrador (Anon, 1993b).

Newfoundland: for 1969–1991, catches of Labrador small and large salmon in Newfoundland were included in total mortalities as the product of the ratio of tags caught in Newfoundland to Labrador and the catch in Labrador. For small salmon the ratio was (24/(100-24)) = 0.32 and for large salmon it was (41/(137-41) = 0.43).

1998-2001

For the years, 1998–2001 when only one or two counting projects took place in Labrador, the raising factors previously used and explained of 1.04 to 1.49 for small salmon and 1.05 to 1.27 for large salmon were used to estimate returns and spawners for Labrador from the overall PFA minus catches in Greenland, as was the case in previous years. However, in this case returns to rivers were derived for Labrador by subtracting landings in food fisheries. Also, catches in 1994–2006 were updated to reflect changes made to catch statistics in Labrador from the Licence Stub Return System. Procedures for the collection and compilation of commercial and angling fishery data are described in Ash and O'Connell (1987) for fishery years 1974–1996. For years 1969–1974, commercial catch data came from Anon. (1978). In 1997, the angling catch statistics were converted to a licence stub system (O'Connell *et al.*, 1998) which continues to present.

2002-present

Counting projects occur on three to four Labrador rivers; out of about 100 extant salmon rivers. Because they occur on the same rivers each year, it is possible to extrapolate from abundance for small and large salmon per accessible drainage areas in these monitored rivers to unsurveyed ones in the remainder of Labrador. The accessible drainage areas were 9267 km² for Lake Melville (SFA 1A), 25 485 km² for Northern Labrador (SFA 1B), 28 160 km² for Southern Labrador (SFA 2), and 2651 km² for the Straits Area (SFA 14B). Accessible drainage area in the counting facility rivers was 1878 km² resulting in an expansion factor of 35 to one. Not all rivers in Lake Melville were included due to a lack of information on presence of salmon populations in rivers in this region of Labrador. Lake Melville rivers whose drainage areas were included are Sebaskachu, Cape Caribou, Goose, MacKenzie, Kenamu, Caroline and Traverspine.

Abundances for SFAs 1A and 1B were derived from English River returns with maximum and minimum values developed using the observed variability of relative abundances in SFA 2. Total returns and spawners for Labrador are estimated by Monte Carlo simulation based on 10 000 random draws from the range of values assuming abundances per km² of accessible drainage were uniformly distributed. The relative abundances (per km²) for each SFA were then multiplied by the total accessible drainage area to derive total returns of small and large salmon. Ranges of values were developed to convert numbers of small and large salmon to numbers of 1SW and 2SW salmon from scale age information collected from counting fences and angling fisheries in Labrador. A bootstrap procedure was used to develop estimates of the proportions of sea age 1 salmon in estimates of large salmon returns and spawners, proportions of sea age 2 salmon in estimates of large salmon returns and returns.

Sea age correction factors were:

Small to 1SW - 96 to 100% Large to 2SW - 60 to 71% Small overlap in large - 12 to 21%

Spawners of 18W and 2SW salmon were derived by subtraction of angling catches including an estimate of hook and release mortalities (10%) from the returns.

2.3.1.2 Newfoundland

Inputs for the run-reconstruction model for Newfoundland include estimates of small, large and 2SW returns and spawners to rivers (minimum and maximum). The methods used to estimate returns and spawners to the rivers in Newfoundland are described by Reddin and Veinott (2010). In brief, returns and spawner estimates were derived from recreational fishery exploitation rates of retained small salmon for rivers with enumeration facilities; and ratios of large to small salmon were utilized to estimate large salmon. Exploitation rates were then applied to all rivers with reported angling catches. A non-parametric bootstrap technique was used, whereby exploitation rates and ratios of large to small salmon from rivers with enumeration facilities were chosen at random with replacement. The 95th confidence interval from 500 iterations of the weighted exploitation rate and ratio of large to small salmon was applied to angling catches on a Salmon Fishing Area (SFA) basis. The midpoint of the 95th confidence interval was used as the minimum and maximum estimate returns of

large and small salmon in each SFA. Estimates of 2SW returns are based on the expected proportion of 2SW in the large salmon category (≥63 cm). Commercial and recreational angling catches were derived as described for Labrador (2.3.1.1). Spawners in all years were determined as the returns to rivers minus angling catches including an adjustment for hook-and-release mortality.

2.3.1.3 Québec

In order to estimate abundance of stocks, rivers were classified into six categories (C1–C6) depending on the information available to estimate salmon returns (according to the method of Caron and Fontaine, 1999), with C1 being the most reliable evaluation and C6 the least. C1 corresponds to a river where the evaluation of the returns is based on a counting method, either from a fence or from a visual count through snorkelling or from a canoe. C2 uses the same evaluation, but without knowing the number of small and large salmon, which is then estimated from proportions reported in the sport fishing landings and, if necessary, the catch and release. Salmon returns on C3 rivers are determined based on multiple correlation factors, using catch number, fishing effort, season duration and river accessibility distance (Guillouët, 1993).

When estimation of the returns using a C1–C3 category is not possible, and when data of returns from previous years are available, the C4 category is used. C4 assumes that interannual variations in salmon returns in the targeted river are approximately the same as variations observed in the other rivers of the corresponding region. Category C5 is for rivers where only landings data are available. In these rivers the salmon run is estimated from the average regional exploitation rate. Finally, a few small rivers have essentially no available data. C6 then assumes that the run is related to the available river salmon habitat and is estimated with respect to rivers of the same area for which run estimates and salmon habitat area are known. Estimated numbers of returns from C4 to C6 cannot be used to assess relative to attainment of conservation limits. However, they provide at least approximate numbers to estimate returns and spawners for salmon rivers in Québec.

The evaluation of the uncertainty associated with return estimates depends on the river category. For C1 and C2 rivers, the correction factor for the minimum and maximum number of returns is $\pm 5\%$ and $\pm 10\%$ for all rivers with a fish ladder and for all others in zones Q1 to Q3 and Q10. The correction factor for rivers with darker water from zones Q5, Q6 and Q7 is rather $\pm 10\%$ and $\pm 30\%$. For the other categories, an uncertainty of $\pm 25\%$ is associated with salmon return estimates, except for category C3 where calculation depends on the method of Guillouët (1993).

The number of spawners is obtained using the return estimate minus all river catches, which include landings and other types of removal. In most cases, river catches include landings from sport fishing only, which may be conducted by native people such as that on the Betsiamites River. The other types of removal are of limited number and include mainly natural mortality, salmon captured for hatchery use and subsistence fishing when practised in river.

Overall return estimates for all Québec rivers are obtained by adding in-river salmon returns, commercial fishing (when operated), native people subsistence fishing when practised in estuaries and an estimate of non-registered landings. However, little scientific data are available on non-registered landings and thus, estimates are based on good judgment, following consultations with regional biologists.

2.3.1.4 Gulf

Estimation of returns and spawners are developed for the four salmon fishing areas of Gulf Region (SFAs 15 to 18).

SFA 15

The major river in this area is the Restigouche River. The returns and spawners are estimated for the Restigouche River exclusive of returns to the Matapedia River, which are included in Québec zone Q1. The Restigouche River stock assessment is based on angling catch with assumed exploitation rates between 30% (min.) and 50% (max) with estuary catches added back after the estimates of returns. Return and spawner estimates for SFA 15 are based on Restigouche River data, scaled up for SFA 15 using angling data. The return and spawner estimates for SFA 15 are derived from the return and spawner estimates for Restigouche (New Brunswick). The minimum and maximum return and spawner estimates are derived from the minimum and maximum ratios of angling catch in all of SFA15 relative to angling catch in Restigouche (New Brunswick) (min = 1.117; max = 1.465). Harvests represent retained angling catch plus 6% catch and release mortality for released fish. The proportion of 2SW in large salmon numbers is based on aged scale samples from angling, trapnets, and broodstock. In the years when no scale samples analysis is available, a mean value of 0.65 is used.

SFA 16

The most important Atlantic salmon river in SFA 16 is the Miramichi River. The Miramichi makes up 91% of total rearing area of SFA 16 and returns to the river are assessed annually. For 1971 to 1991, minimum and maximum values are based on capture efficiencies of the Millbank estuary trapnet representing a lower CI of -20% of the estimate and upper CI of 33% of the estimate. For 1992 to 1997, minimum and maximum are lower and upper CI and based on estimate bounds of -18.5% to +18.5%. Since 1998 to the present, minimum and maximum are 5th and 95th percentile range from a Bayesian hierarchical model used in the assessment. Returns to SFA 16 are Miramichi returns (Minimum, Maximum) / 0.91. Proportion 1SW in small salmon is from scale ageing; proportions have varied from 0.97 to 1.0. Proportion 2SW in the large salmon category is obtained from scale ageing. Spawners are returns minus harvests. For 1998 to 2011, the harvest of large salmon is estimated as the sum of the aboriginal fisheries harvests for large salmon and 1% of the large salmon catch (30% exploitation rate, 3% catch and release mortality). The harvest of small salmon is estimated as 30% of the small salmon return plus the harvest from the aboriginal fisheries.

SFA 17

For 1970–1994, small returns are estimated from retained small salmon catch in the Morell River divided by the river-specific exploitation rate. Salmon catch in the Morell River was estimated in 1970–1990 by DFO Fisheries Officers; and in 1991, 1992, and 1994 by angler mail-out surveys. The number of small retained salmon in 1993 was not recorded, so the number used is the mean for 1986–1992. For 1970–1993, exploitation rate was taken as the mean of exploitation rates estimated for 1994, 1995, and 1996 (0.317). For 1994, exploitation rate was 0.34. The min and max of small returns are calculated using exploitation +/- 0.1; e.g. 0.34 +/- 0.1 gives 0.24 and 0.44. Large returns = (number of small returns/proportion small) - number of small returns. For 1970–1980, proportion small is calculated from numbers of small and

large salmon in the angling catch of each year. For 1981–1994, proportion small is taken from counts at the Leards Pond trap on the Morell River. Small spawners = number of small recruits-number of small retained. Large spawners = number of large recruits - number of large retained.

Spawner estimates for 1995 to the present are derived from redd counts in 23 rivers. For years and rivers in which redd counts are unavailable, redd numbers are estimated by linear interpolation from the preceding and succeeding count year. Redd numbers in years prior to the first count are taken as the first count. Redd numbers in years after the last count are taken as the last count. Female spawners are estimated from the ratio of 3.357 redds/female spawner, measured in the West River in 1990. Total spawners are estimated from size-specific sex ratios derived from counts at Leards and Mooneys Ponds, Morell River, in 1986-2001. The proportion of salmon that are large is taken as 0.5 in the Cains, Carruthers, Trout (Coleman), Morell, Cardigan, West, and Dunk Rivers, and 0.9 in all other rivers. Spawners are presented as Min (estimated spawners -20%) and Max (estimated spawners + 20%). Returns are spawners + total estimated fishing mortality, including angler catches, hook-and-release mortality, and native harvests. Angler catches and hook-andrelease mortality are estimated from angler card surveys. Returns are presented as Min (estimated returns -20%) and Max (estimated returns + 20%). It is assumed that large salmon and 2SW salmon are equivalent.

SFA 18

Returns and spawners to SFA 18 are derived from estimates of returns and spawners to the Margaree River, adjusted for the ratio of the SFA 18 angling catch to the Margaree River catch. For small salmon, the ration of SFA 18 catch to Margaree catch varies between 1.15 and 2.71 for years 1984 to 2004. For large salmon, the ratio of SFA 18 catch to Margaree catch varies between 1.08 and 2.32 for years 1984 to 2004. Returns to Margaree River are estimated using various techniques.

- 1970 to 1983 angling catch divided by range of exploitation rates with maximum exploitation rate of 0.37 and minimum exploitation rate of 0.215;
- 1984 to 1986 based on annual assessments;
- 1987 to present angling catch and effort data from logbooks and provincial licence stubs are used to derive the returns. The catchability coefficient per rod day is estimated from angling catch and effort data for the years 1988 to 1996 when mark and recapture programmes were used to estimate returns, independently from angling data.

Spawners for 1970–1983 equal returns minus removals. Spawners for 1984 to the present equal returns minus catch for small salmon and returns minus catch, corrected for 5% mortality, for large salmon. 2SW salmon represent between 0.77 and 0.87 of large salmon returns and spawners.

2.3.1.5 Scotia-Fundy

Salmon originating in rivers of the Atlantic coast of Nova Scotia and southwest New Brunswick in Salmon Fishing Areas (SFAs) 19–21 and the portion of SFA 23 outside the inner Bay of Fundy comprise the Scotia-Fundy stocks. With the exception of at least one stock in SFA 19, they have a large salmon component that migrates to the North Atlantic/Labrador Sea (Amiro *et al.*, 2008). Estimates of returns and spawning escapement for the Scotia-Fundy stocks are provided as inputs to the run-

reconstruction model. Inner Bay of Fundy Atlantic Salmon (SFA 22 and part of SFA 23) have been federally listed as endangered under the Canadian Species at Risk Act and are not included as inputs into the run-reconstruction model. With the exception of one population, inner Bay of Fundy stocks have a localized migration strategy while at sea and an incidence of maturity after one winter at sea.

Consistent with the requirements of the model, a range (minimum to maximum) of returns and spawning escapement for the Scotia-Fundy stocks is provided for the run-reconstruction model. The methods used to estimate total returns and spawners are described by Amiro et al. (2008). In brief, for SFAs 19-21, the escapement is based on the count of small and large salmon at the Morgan Falls fish-way on the LaHave River from 1970 to the present year, scaled up to the region using the relationship between this count and the recreational catch data for rivers in SFA 19 to 21 from 1970 to 1997 and a catch rate for the LaHave River from 1970 to 1997. Estimates of the returns also include estimates of landings in the commercial salmon fisheries in SFA 19-21 from 1970 to 1983. The model is fitted using maximum likelihood, and the 90% confidence limits are carried forward as the minimum and maximum values. In SFA 23 from 1970 until 1992, estimates of total 1SW and large wild-origin salmon returns are based on the estimated number of returns destined for tributaries above Mactaquac Dam on the Saint John River; this includes in-river and outer-Fundy commercial landings (1970-1971 and 1981-1983), in river aboriginal harvests (since 1974), and counts at Mactaquac Dam. These estimates are raised by the proportion of the total accessible productive habitat in SFA 23 that is upstream of Mactaquac Dam (0.4-0.6). Hatchery-origin returns were attributed to above Mactaquac Dam only and no hatchery 1SW and MSW returns were estimated for other rivers within SFA 23 (outer Fundy). Since 1993 the estimates of 1SW and MSW returns to the Nashwaak River have been used to estimate the wild production from tributaries of the Saint John River below Mactaquac Dam. The estimated 1SW and MSW returns to the Nashwaak River (above Counting Fence), is raised by the proportion of the total production area accounted for below Mactaquac (0.21-0.3) and then added to the above Mactaquac totals.

2.3.1.6 USA

Total salmon returns and spawners for USA rivers are based on trap and weir catches and for the small rivers in Maine that do not have fish counting facilities, estimates of spawners were based on redd counts.

2.3.2 Improvements to NAC input data

Modifications to input variables used in assessments for the NAC area are reported by WGNAS in the year in which they are first implemented.

Over recent years, efforts have been made to reduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes). However, the methods used to derive estimates of unreported catch vary markedly between countries. For example, some countries include only illegally caught fish in the unreported catch, while other countries include estimates of unreported catch by legal gear as well as illegal catches in their estimates.

Descriptions of the national approaches used for evaluating unreported catches have been reported at various WGNAS meetings (e.g. ICES, 1996; 2000; 2002; 2010a). In addition, detailed reports describing national procedures for evaluating illegal and

unreported catch, and efforts to minimise this, were submitted by parties to NASCO in 2007 in support of a special theme session on this issue. Full details are available at: http://www.nasco.int/pdf/2007%20papers/CNL(07)26.pdf

Input data commonly rely on rod catches and the practice of catch-and-release has become increasingly important in recent years to reduce levels of exploitation on stocks. As the practice is increasing, WGNAS have previously recommended (ICES, 2010a) that consideration should be given to incorporating mortality associated with this practice in river-specific, regional and national assessments.

The procedures currently used to incorporate catch-and-release and unreported catches into regional, national and international assessments are summarised at Appendix 1 (from ICES, 2010a).

2.4 Biological and other data requirements

As noted previously, many of the 'conventional' data requirements (e.g. marine survey data and commercial cpue) used in the assessment of other commercially important fish species are inappropriate to salmon. A range of biological, catch and exploitation rates and other data pertinent to appropriate stock assessments are however, collected and made available to WGNAS to help inform assessments and to aid in responding to the various questions posed by NASCO.

Appendix 2 of this Stock Annex provides an overview of current and possible future data requirements for Atlantic salmon assessment/ scientific advice. This was compiled at a recent meeting of WGNAS (ICES, 2013) in relation to monitoring requirements under the European Data Collection Framework (DCF) and following a more detailed review of the data requirements under DCF (ICES, 2012c). This table illustrates the type of information collected/available, but is provided for illustrative purposes only. It should be noted that many Atlantic salmon producing countries fall outside the DCF provisions, which only relate to countries within the European Union. Further, Sovereign states are responsible for the regulation of salmon fisheries within their areas of jurisdiction. Formal ICES catch advice is only required for the distant water salmon fisheries, which take salmon originating in rivers of another party.

3 Assessment methods

In managing Atlantic salmon fisheries, NASCO has adopted a fixed escapement strategy (Potter, 2001), in recognition of the importance of the spawning stock to subsequent recruitment. Therefore, in managing the distant water fisheries at Faroes and West Greenland, the spawning requirements of the rivers contributing to these fisheries must be defined. Management advice, expressed as allowable harvest (tonnes), is then predicated on a forecast of salmon abundance prior to the fishery such that the spawning requirements of the contributing stocks can be achieved. The provision of catch advice thus proceeds through a number of steps:

The definition of spawning objectives;

The development of a measure of abundance prior to the fishery; i.e. the prefishery abundance or PFA;

A measure of the spawning stock contributing to the PFA;

A model to forecast the PFA;

The development of a risk analysis framework for the catch advice.

These steps are described in detail in the following sections, subdivided as necessary for the different distant water fisheries and the various stock complexes which contribute to the two fisheries (Greenland and Faroes).

3.1 Definition of spawning objectives

3.1.1 Management objectives and reference points

Conservation limits (CLs) for North Atlantic salmon have been defined by ICES as the stock level that will achieve long-term average maximum sustainable yield (MSY). NASCO has adopted the following definition of CLs (NASCO, 1998): 'The CL is a limit reference point; having populations fall below these limits should be avoided with high probability.'

Atlantic salmon has characteristics of short-lived fish stocks; mature abundance is sensitive to annual recruitment because there are only a few age groups in the adult spawning stock. Incoming recruitment is often the main component of the fishable stock. For such fish stocks, the ICES MSY approach is aimed at achieving a target escapement (MSY Bescapement, the amount of biomass left to spawn). No catch should be allowed unless there is a high probabaility that this escapement can be achieved. The escapement level should be set so there is a low risk of future recruitment being impaired, similar to the basis for estimating Bpa in the precautionary approach. In short-lived stocks, where most of the annual surplus production is from recruitment (not growth), MSY Bescapement and Bpa might be expected to be similar.

It should be noted that this is equivalent to the ICES precautionary target reference points (S_{pa}) . Therefore, stocks are regarded by ICES as being at full reproductive capacity only if they are above the precautionary target reference point. This approach parallels the use of precautionary reference points used for the provision of catch advice for other fish stocks in the ICES area.

For the assessment of the status of stocks and advice on management of national components and geographical groupings of the stock complexes, where there are no specific management objectives:

- ICES requires that the lower bound of the 90% confidence interval of the current estimate of spawners is above the CL for the stock to be considered at full reproductive capacity (equivalent to a probability of at least 95% of meeting the CL).
- When the lower bound of the confidence limit is below the CL, but the midpoint is above, then ICES considers the stock to be at risk of suffering reduced reproductive capacity.
- Finally, when the midpoint is below the CL, ICES considers the stock to be suffering reduced reproductive capacity.

Ideally, Atlantic salmon should be assessed and managed on the basis of river-specific stock units, the scale corresponding best to the spawner to recruitment dynamic (Chaput, 2012). In reality, this is not the case for the majority of rivers, although efforts are continuing to develop river-specific CLs and assessment protocols and developments are reported annually to WGNAS (e.g. ICES, 2013).

The risk assessment frameworks applied by WGNAS directly evaluate the risk of meeting or exceeding the stock complex objectives. Managers can choose the risk level which they consider appropriate. ICES considers however that to be consistent

with the MSY and the precautionary approach, and given that the CLs are considered to be limit reference points and to be avoided with a high probability, then managers should choose a risk level that results in a low chance of failing to meet the CLs. ICES recommends that the probability of meeting or exceeding CLs for individual stocks should be greater than 95% (ICES, 2012b).

3.1.2 Reference points in the NEAC area

River-specific CLs have been derived for salmon stocks in some countries in the NEAC area (France, Ireland, UK (England & Wales) and Norway). An interim approach has been developed for estimating national CLs for countries that cannot provide one based upon river-specific estimates. This approach is based on the establishment of pseudo stock–recruitment relationships for national salmon stocks (Potter *et al.*, 2004).

The NEAC-PFA run reconstruction model (see below) provides a means of relating estimates of the numbers of recruits to the numbers of spawners. The numbers of 1SW and MSW spawners are converted into numbers of eggs deposited using the proportion of female fish in each age class and the average number of eggs produced per female. The egg deposition in year 'n' is assumed to contribute to the recruitment in years "n+3" to "n+8" in proportion to the numbers of smolts produced of ages 1 to 6 years respectively. These proportions are then used to estimate the 'lagged egg deposition' contributing to the recruitment of maturing and non-maturing 1SW fish in the appropriate years. The plots of lagged eggs (stock) against the 1SW adults in the sea (recruits) are presented as 'pseudo stock-recruitment' relationships for each homewater country or region that is unable to provide river-specific CLs. In countries where with more than one region, the analysis is carried out for each region separately and the resulting estimates are summed to provide a national figure.

As noted previously, ICES currently define the CL for salmon as the stock size that will result in the maximum sustainable yield (MSY) in the long term. However, it is not straightforward to estimate this point on the stock-recruitment relationships established by the national PFA run-reconstruction models, as the replacement line (i.e. the line on which 'stock' equals 'recruits') is not known for these relationships. This is because the stock is expressed as eggs, while the recruits are expressed as adult salmon. To address this, WGNAS has developed a method for setting biological reference points from the national/regional pseudo stock-recruitment datasets (ICES, 2001). This model assumes that there is a critical spawning stock level below which recruitment decreases linearly towards zero and above which recruitment remains constant. The position of this critical stock level is determined by searching for the stock value that provides the line of best fit for the stock and recruitment data provided by the PFA run-reconstruction model as determined by the residual sum of squares. This point is a proxy for Slim and is therefore defined as the CL for the stock, and is indicated by the inflection point in the hockey-stick relationship (e.g. see example at Figure 3.1.2.1).

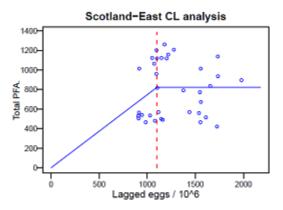


Figure 3.1.2.1. Pseudo stock-recruitment relationship for UK (Scotland) eastern region (from ICES, 2013).

Where river-specific estimates of CLs have been derived for all the rivers in a country or region, these are aggregated to provide national estimates. For countries where the development of river-specific CLs has not been completed, the method described above has been used (see example in Table 3.1.2.1, from ICES, 2013). The estimated national CLs are then summed to provide aggregate CLs for the northern and southern NEAC stock complexes (Table 3.1.2.1).

The CLs have also been used to estimate the spawning escapement reserves (SERs). These represent the CLs increased to take account of natural mortality between the recruitment date, 1st January, and the return to homewaters for maturing and non-maturing 1SW salmon from the northern NEAC and southern NEAC stock complexes (Table 3.1.2.1).

Table 3.1.2.1. Conservation limits (CLs) for NEAC countries and stock complexes estimated from river-specific values, where available, or the national PFA run-reconstruction model. Spawner escapement reserves (SERs) are also included for each stock complex.

	National	Model CLs	River Specific CLs		Conservati	on limit used	SER		
	1SW	MSW	1SW	MSW	1SW	MSW	1SW	MSW	
Northern Europe									
Finland	18,755	13,819			18,755	13,819	22,819	23,788	
Iceland (north & east)	6,032	1,620			6,032	1,620	7,450	2,788	
Norway			63,939	72,198	63,939	72,198	81,397	120,589	
Russia	67,710	38,913			67,710	38,913	86,086	70,285	
Sweden	1,181	1,196			1,181	1,196	1,527	2,090	
			Stock Cor	nplex	157,617	127,745	199,279	219,540	
	National	Model CLs	River S	pecific CLs	Conservati	on limit used	S	ER	
	1SW	MSW	1SW	MSW	1SW	MSW	1SW	MSW	
Southern Europe									
France			17.400	5,100	17,400	5,100	22,488	9,467	
Iceland (south & west) 17,751	1,158	, ,	,	17,751	1,158	21,926	1,994	
Ireland			211,471	46,943	211,471	46,943	269,210	78,407	
UK (E&W)			54,812	30,203	54,812	30,203	69,778	51,993	
UK (NI)			21,649	2,437	21,649	2,437	26,553	4,100	
UK (Sco)	245,912	187,518		·	245,912	187,518	313,054	319,685	
			Stock con	nplex	568,995	273,360	723,008	465,646	

WGNAS considers the current CL and SER levels may be less appropriate to evaluating the historical status of stocks (e.g. pre-1985), that in many cases have been estimated with less precision.

3.1.3 Reference points in the NAC area

In many regions of North America, the CLs are calculated as the number of spawners required to fully seed the wetted area of the river. The methods and values used to derive the egg and spawner conservation requirements for Atlantic Canada are documented in O'Connell et al. (1997). CLs have generally been derived using freshwater production dynamics translated to adult returns to estimate the spawning stock for maximum sustainable yield (MSY). Data were available for a limited number of stocks and these values were transported to the remaining rivers using information on habitat area and the age composition of the spawners. A similar procedure was used to determine the CLs for rivers in the USA (ICES, 1995). In Québec, adult-to-adult stock-recruitment relationships for six rivers were used to define the CLs for the other rivers (Caron et al., 1999). The definition of conservation in Canada varies by region and in some areas, historically the values used were equivalent to maximizing /optimizing freshwater production. These are used in Canada as limit reference points and they do not correspond to MSY values. Reference points for Atlantic salmon are currently being reviewed for conformity with the Precautionary Approach policy in Canada. Revised reference points are expected to be developed.

The NAC conservation requirements for 2SW salmon (only these are required in developing catch options for the West Greenland fishery) are summarised in Table 3.1.3.1 (from ICES, 2013). These are calculated from the adult age structure within the different regions and total 123 349 2SW salmon for Canada and 29 199 2SW salmon for the USA, for a combined total of 152 548.

Table 3.1.3.1. 2SW Conservation limits (CLs) for the six regions in the NAC area estimated from river-specific values.

COUNTRY AND COMISSION AREA	STOCK AREA	2SW SPAWNER REQUIREMENT
	Labrador	34 746
	Newfoundland	4022
	Gulf of St Lawrence	30 430
	Québec	29 446
	Scotia-Fundy	24 705
Canada Total		123 349
USA		29 199
North American Total		152 548

3.2 Estimating PFA

Estimates of PFA are derived by run-reconstruction methods. These work back in time from estimates of abundance in homewaters to earlier periods of the salmon's life cycle by adding in catches at appropriate times and adjusting for survival. The run-reconstruction approach was first presented at ICES in 1992 and was subsequently adopted for stocks on both sides of the Atlantic (Rago *et al.*, 1993a; Potter and Dunkley, 1993; Potter *et al.*, 1998; 2004). The main advantage of backwards-running, run-reconstruction models over alternative forward-running approaches is that more extensive data are available on adult returns (e.g. traps, counters and catch data) than on freshwater production of juveniles. In addition, rates of natural mortality (M) were thought to be lower and more stable for large

salmon after their first winter in the sea than during the post-smolt phase (Potter *et al.*, 2003).

The models used to estimate PFA take the generalised form:

$$PFA = Nh * \exp(Mt_h) + \sum_{i} C_i * \exp(Mt_i)$$

Where: Nh is the number of adult fish returning to homewaters, C_i the catch of fish from the stock in each interception fishery i (operating before the fish return to homewaters), M the monthly instantaneous rate of natural mortality of salmon in the sea after the first sea-winter, t_i the time in months between the PFA date and the midpoint of fishery i, and t_h is the time in months between the PFA date and the midpoint of the return of fish to homewaters. Coastal catches are also added to the estimate where appropriate.

3.2.1 NEAC area run reconstruction model

The original model used to estimate the PFA of salmon from countries in the NEAC area was described by Potter *et al.* (2004); modifications have been described in subsequent WGNAS reports. PFA in the NEAC area is defined as the number of 1SW recruits on January 1st in the first sea winter. As there are relatively few fish of sea age three or more in most stocks, the model caters for two age groups, 1SW and MSW, the latter including all fish of sea age two or more that are treated as a single cohort. The model is therefore based on the annual catch in numbers of 1SW and MSW salmon in each country. These are raised to take account of minimum and maximum estimates of non-reported catches and exploitation rates of these two sea-age groups.

Thus, for each country (or region) c in year y, the total number of fish of sea age a caught in homewater fisheries ($Ch_{a,y,c}$) is calculated by dividing the declared catch ($Cd_{a,y,c}$) by the non-reporting rate (1 - $U_{a,y,c}$):

$$Ch_{a,y,c} = Cd_{a,y,c} / (1 - U_{a,y,c})$$

where: $\dot{U}_{a,y,c}$ is the estimated proportion of the total catch that is unreported or discarded. The number of fish returning to homewaters (Nh_{a,y,c}) is estimated by dividing the total homewater catch by the exploitation rate (H_{a,y,c}):

$$Nh_{a,y,c} = Ch_{a,y,c} / H_{a,y,c}$$

As the model provides estimates of total returns and total catch (including non-catch fishing mortality), it is then also possible to estimate the spawner escapement (Nsa,y,c):

$$Ns_{a,y,c} = Nh_{a,y,c} - Ch_{a,y,c}$$

Total catches in the Faroese (Cf_{a,y}) and West Greenland (Cg_{a,y}) fisheries are similarly calculated by correcting the declared catches for non-reporting, but they are not raised for the exploitation rate, because the uncaught fish are accounted for from the returns to homewaters. The West Greenland fishery only exploits salmon that would otherwise mature as MSW fish, although the majority are 1SW fish in the summer that they are caught; for the purpose of the model, all are classed as 1SW. The Faroese fishery exploits predominantly MSW salmon, but also a small number of 1SW fish, 78% of which have been estimated to be maturing (ICES, 1994). Over the past two

decades, a substantial proportion of the fish caught in the Faroese fishery have been escapees from salmon farms, and these are discounted from the assessment of wild stocks on the basis of data from Hansen *et al.* (1999). The incidence of farm escapees in the West Greenland catch is thought to be <1.5% (Hansen *et al.*, 1997), so this portion is ignored in the model. The total estimated catches of wild fish in both distant-water fisheries are assigned to the PFA for different countries on the basis of historic tagging studies (Potter, 1996).

The returns to homewaters and catches in the distant-water fisheries of 1SW and MSW salmon are then raised to take account of the marine mortality between January 1st in the first sea winter (the PFA date) and the mid-point of the period over which the respective national fisheries operate. WGNAS determined a natural mortality value of 0.03 per month to be the most appropriate (ICES, 2002) and a range 0.02 to 0.04 is applied within the model in a Monte Carlo simulation. Thus, the PFA of maturing 1SW fish (PFAm), survivors of which will return to homewaters as 1SW adults, is:

$$PFAm_{y,c} = Nh_{1,y,c} * exp(Mt_{h,1,c}) + 0.78 * Cf_{1,y} * w_y * pf_{1,c} * exp(Mt_{f,1,c})$$

and the PFA of non-maturing 1SW fish (PFAn), survivors of which will return to homewaters as MSW adults, is:

$$\begin{split} & PFAn_{y,c} = Nh_{2,y+1,c} * exp(Mt_{h,2,c}) + Cg_{1,y} * pg_{1,c} * exp(Mt_{g,1,c}) \\ & + 0.22 * Cf_{1,y} * w_y * pf_{1,c} * exp(Mt_{f,1,c}) + Cf_{2,y+1} * w_{y+1} * pf_{2,c} * exp(Mt_{f,2,c}) \end{split}$$

where indices y and c represent year and country/region, indices 1 and 2 the 1SW and MSW sea age groups, w is the proportion of the Faroese catch that is of wild origin, pf and pg are the proportion of the catches in the Faroese and West Greenland fisheries originating in each country (as indexed), and t_h , t_f and t_g are the times in months between the PFA date and the midpoints of the homewater fisheries, the Faroese fishery, and the West Greenland fishery, respectively, for the year classes and country/region as indexed.

Total 1SW recruitment for the NEAC area in year y is therefore the sum of the maturing 1SW and non-maturing 1SW recruitments for that year for all countries:

$$PFA_y = \sum_{c} PFAm_{y,c} + \sum_{c} PFAn_{y,c}$$

The non-reporting rates, exploitation rates, natural mortality, and migration times in the above equations cannot be estimated precisely, so national experts provide minimum and maximum values based upon best available knowledge that are considered likely to be centred on the true values (ICES, 2003). A Monte Carlo simulation (12 000 trials) is used to estimate confidence intervals on the stock estimates.

Where appropriate to the provision of management advice, the national outputs from the model are combined into stock complexes, such as those for southern and northern NEAC (ICES, 2002). The confidence limits for these combined estimates are derived from the sum of the national variances obtained from the MCS (the covariances are assumed to be small). This model has provided time-series of PFA estimates for NEAC salmon stocks from 1971 to the present.

The model was initially run using 'Crystal Ball' (CB) in Excel (Decisioneering, 1996). However, an updated version of the model which runs in the 'R' programming language (R Development Core Team, 2007) was developed in 2011 (ICES, 2011). This provided a more flexible platform for the further development of the model and to allow its integration with the Bayesian forecast model for the development of catch options (see below). In 2012, the outputs of the CB and 'R' models were compared to examine the approaches taken and to validate the outputs (ICES, 2012a). Since 2013, the run-reconstruction analysis has been completed by WGNAS using the 'R' programme (ICES, 2013). This has also enabled additional sources of uncertainty to be incorporated into the modelling approach (ICES, 2013).

The full set of country-specific data inputs, as used in the most recent assessment (ICES, 2015) is provided at Appendix 3. The 'R' code used for running the model and the additional data input file required to run the model, are available on the ICES WGNAS SharePoint site.

3.2.2 NAC area run reconstruction model

The run–reconstruction model developed by Rago *et al.* (1993a) and described in previous WGNAS reports (ICES, 2008; 2009a) and in the primary literature (Chaput *et al.*, 2005) is used to estimate returns and spawners by size (small salmon, large salmon) and sea age group (2SW salmon) to the six geographic regions of NAC from 1971 to the present. The model takes the form:

$$PFA_{year(i)} = [NR2_{year(i+1)} * e^{MX1} + NC2_{year(i+1)}] * e^{MX10} + NC1_{year(i)} + NG1_{year(i)} + N$$

where: NR2_{year(i+1)} is the sum of 2SW returns to six regions of North America in year i + 1, NC2_{year(i+1)} is the catch of 2SW salmon in Newfoundland and Labrador commercial fisheries in year i + 1, NC1_{year(i)} is the catch of 1SW non-maturing salmon in Newfoundland and Labrador commercial fisheries in year i, NG1_{year(i)} is the catch of 1SW non-maturing salmon of North American origin in the Greenland fishery in year i, and M is the monthly instantaneous natural mortality of 0.03.

The reconstruction begins with the estimation of returns of 2SW salmon in year i + 1 to six regions in eastern North America: Labrador, Newfoundland, Québec, Gulf, Scotia-Fundy, and USA. For the four southern regions, the regional returns include the harvest in the coastal commercial fisheries but this is not the case for Newfoundland and Labrador. For Labrador, the returns to rivers are estimated from the commercial harvest factored by an exploitation rate. The harvest of 2SW salmon in the Newfoundland and Labrador mixed-stock fisheries in year i + 1 is added to the sum of the returns to the six regions (prorated backward for one month of natural mortality - equates to 1st June of year i + 1) to produce the returns to North America. Finally, the harvests of North American origin salmon in the Greenland fisheries in year i and the harvest of non-maturing 1SW salmon in the Newfoundland and Labrador commercial fisheries in year i are added to the prorated returns to North America (ten months between abundance at Greenland on 1st August year i and North America on 1st June year i + 1) to produce the pre-fishery abundance of nonmaturing 1SW salmon of North American origin. An instantaneous natural mortality rate of 0.03 per month is assumed for salmon in the second year at sea for all years (ICES, 2002). Adjustments to the input data resulting from reductions and subsequent closures of commercial fisheries in North America are summarized by Friedland et al. (2003).

Following earlier WGNAS recommendations (ICES, 2008), the run–reconstruction model since 2009 has been developed using Monte Carlo simulation (OpenBUGS; http://mathstat.helsinki.fi/openbugs/; Lunn *et al.*, 2000). This is similar to the approach applied for the NEAC area.

The PFA of the non-maturing component of 1SW fish, destined to be 2SW returns (excluding 3SW and previous spawners) is the estimated number of salmon in the North Atlantic on August 1st of the second summer at sea. As this requires estimates of 2SW returns to rivers, there is always a lag in providing this figure (PFA estimates for year n require 2SW returns to rivers in North America in year n + 1).

The full set of data inputs, as used in the most recent assessment (ICES, 2015) is provided at Appendix 4. The 'R' code used for running the model is available on the WGNAS SharePoint site.

3.2.3 Instantaneous natural mortality rate (M)

The natural mortality rate for salmon after they recruit to the distant water fisheries has been the subject of much discussion. WGNAS originally used a value of 0.01 per month, based upon Doubleday *et al.* (1979), but this was modified to 0.03 per month following a detailed review as part of the EU SALMODEL project (Crozier *et al.*, 2003; ICES, 2002) on the basis of inverse-weight and maturity-schedule models. The rate is assumed to have been constant over the time-series. While mortality may be expected to vary among years and may also be different for maturing and non-maturing 1SW recruits, WGNAS has not had data on which to base the use of different values, or values that change over time. The assumption is, therefore, that the mortality of adult fish after the first winter at sea has not changed and that all the variability of marine mortality has occurred at the post-smolt stage. Efforts are continuing to explore levels of mortality and to better partition this between different stages of the life cycle. The issue was also subject to further investigation within the EU ECOKNOWS project and Bayesian modelling may provide alternative approaches in future.

3.3 PFA forecast models

3.3.1 Introduction

The provision of quantitative catch advice for the distant water fisheries requires estimates of abundance before the fisheries take place. While there has been some use of in-season surveys in the management of these fisheries (NASCO, 2001), such methods are considered too impractical and costly to implement on a widespread scale (Potter *et al.*, 2004). Models have therefore been developed by WGNAS which relate abundance estimates obtained at other life stages to the PFA. The objective has been to account for this relationship in terms of biological or environmental factors that affect natural mortality, and to use this to forecast future stock levels.

An initial PFA forecast model for North American stocks (Rago *et al.*, 1993b) utilised indices of thermal habitat in relation to historically observed PFA (from the runreconstruction model) to predict future PFA. Similar approaches were explored by Crozier *et al.* (2003) for NEAC stocks. However, while statistically significant temperature indices could be constructed, the relationships were not always consistent or intuitively correct. Alternative approaches were therefore explored for NEAC; these are described by Potter *et al.* (2004). More recently work by the ICES Study Group on Salmon Stock Assessment and Forecasting (SGSSAFE) has, however,

resulted in the development of Bayesian forecast models for both NAC and NEAC (ICES, 2009a; 2011; Chaput, 2012).

In the latest models, PFA dynamics by complex are modelled using the estimates of adult spawners, adjusted to the number of eggs per fish based on life-history characteristics of the age groups within each region of the stock complexes (ICES, 2011; Chaput, 2012). The spawner to PFA dynamic is modelled as:

$$PFA_{v} = e^{\alpha_{y}} LE_{v} e^{\varepsilon}$$

where: α_y is the productivity parameter from eggs (×1000) to PFA (number of fish) for PFA year y (on a log-scale), LE $_y$ the estimated lagged eggs (×1000) corresponding to the PFA cohort in year y, and the progress of α_y is modelled as $a_{y+1}=a_y+\varepsilon$, with ε ~ $N(0,\sigma^2)$.

Productivity is modelled as an integration of survival in freshwater and during the first year at sea. An important assumption is the absence of heritability of age at maturity, i.e. all eggs are considered equivalent regardless of the age of the spawners. Lagged eggs refer to the adjustment of the egg depositions to correspond to the expected age at smoltification. Spawners in year 'n' contribute to recruitment in years 'n+3' to 'n+8' depending upon the relative proportions of one to six year-old smolts that they produce. For example, spawners in year 'n' produce eggs that hatch in year 'n+1' and may produce one year-old smolts in year 'n+2', which would become 1SW recruits in year 'n+3'. Any two year-old smolts from the same spawners would produce 1SW recruits in year 'n+4', etc.

At the stock complex level, lagged eggs are the sum of the eggs from the spawners in year y - (s + 2) weighted by the proportion of the smolts produced at age s in region k summed over regions in the complex. Two years are added to the smolt age, for the spawning year and smolt migration year, to lag the eggs to the corresponding year of PFA:

$$LE_y = \sum_{k} \sum_{s} Eggs_{y-(s+2),k} * prop_{s,k}$$

3.3.2 NEAC PFA Forecast model

A forecast model to estimate PFA for all four NEAC stock complexes has been developed in a Bayesian framework by the Study Group on Salmon Stock Assessment and Forecasting (SGSSAFE). The model was originally reported to WGNAS in 2009 (ICES, 2009a), but was subsequently refined and has been in use by WGNAS in its present form since 2011 (ICES, 2011). The models for the northern and southern NEAC stock complexes have exactly the same structure and are run independently. A Directed Acyclical Graph (DAG) for the models is provided in Figure 3.3.2.1. The model considers both the maturing *PFA* (denoted *PFAm*) and the non-maturing PFA (denoted *PFAnm*). The full code used for running the model is available on the WGNAS SharePoint site.

A disaggregated version of the Bayesian model has since been developed using the same structure to provide forecasts at a country level, for all countries in both southern and northern NEAC model implementations (ICES, 2013). In these, countries are linked hierarchically only through the variance on the productivity parameter "a". There is no modelling linkage between the northern and southern complexes.

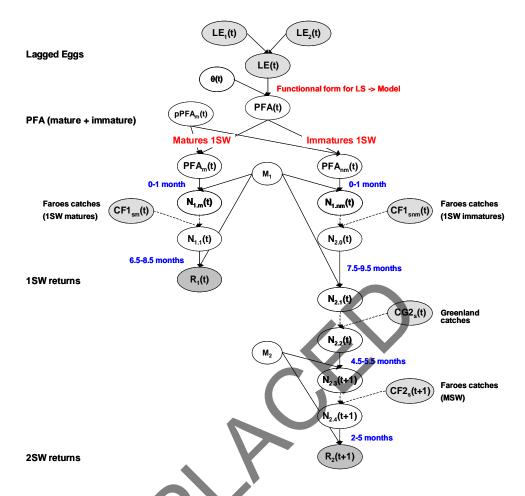


Figure 3.3.2.1. Directed Acyclical Graph (DAG) of the structure of the combined sea age model for the southern NEAC and northern NEAC forecast models. Ellipses in grey are observations (or pseudo-observations) derived from sampling programmes or from submodels (run-reconstruction).

The PFA is modelled using the summation of lagged eggs from 1SW and MSW fish (LE) for each year t and an exponential productivity parameter (a).

$$PFAt = LEt^* exp(at)$$

The productivity parameter (a) is the proportionality coefficient between lagged eggs and PFA. This is forecasted one year at a time (a_{i-1}) in an auto correlated random walk, using the previous year's value (a) as the mean value in a normal distribution, with a common variance for the time-series of a.

$$at+1=at+\varepsilon$$
, $\varepsilon \sim N(0, a.\sigma 2)$

The maturing *PFA* (denoted *PFAm*) and the non-maturing *PFA* (denoted *PFAnm*) recruitment streams are subsequently calculated from the proportion of *PFA* maturing (*p.PFAm*) for each year *t. p.PFAm* is forecast as an autocorrelated value from a normal distribution based on a logit scale, using the previous year's value as the mean and a common variance across the time-series of *p.PFAm*.

```
logit.p.PFAmt+1 \sim N(logit.p.PFAmt, p.\sigma2)
logit.p.PFAmt = logit(p.PFAmt)
```

Uncertainties in the lagged eggs were accounted for by assuming that the lagged eggs of 1SW and MSW fish were normally distributed with means and standard deviations derived from the Monte-Carlo run-reconstruction at the scale of the stock complex. The uncertainties in the maturing and non-maturing PFA returns are derived in the Bayesian forecast models through the pseudo-observation method proposed by Michielsens *et al.* (2008), as used in the NAC model.

The natural mortality in the post-PFA time point was assumed constant among years, centred on an instantaneous rate value of 0.03 per month with a 95% confidence interval range of 0.02 to 0.04.

Catches of salmon at sea in the West Greenland fishery (as 1SW non-maturing salmon) and at Faroes (as 1SW maturing and MSW salmon) were introduced as covariates and incorporated directly within the inference and forecast structure of the model. For southern NEAC, the data are available for a time-series of lagged eggs and returns commencing in 1978. Although the return estimates to southern NEAC begin in 1971, the lagged eggs are only available from 1978 due to the smolt age distributions (one to five years). For northern NEAC, data are available for a shorter time-series. Return and spawner estimates begin in 1983, but due to the smolt age distributions (one to six years), the lagged eggs are only available from 1991 onward. The models are fitted and forecasts derived in a consistent Bayesian framework.

The model provides forecasts for maturing and non-maturing stocks for both southern and northern NEAC complexes (and countries) for five years. Risks are defined each year as the posterior probability that the PFA would be greater than or equal to the age and stock complex/ country specific Spawner Escapement Reserves (SERs), under the scenario of no exploitation.

The country disaggregated version of the Bayesian NEAC inference and forecast model incorporates country specific catch proportions at Faroes, lagged eggs and returns of maturing and non-maturing components. Model structure and operation is as described above, incorporating country and year indexing. Linkage between countries in the model is through the common variance parameter associated with the productivity parameter (a) (the proportionality coefficient between lagged eggs and PFA), which is forecast forward and used along with the forecast proportion maturing to estimate the future maturing and non-maturing PFAs. The evolution of a is independent between countries with the exception of its associated variance. Evolution of the proportion maturing (p.PFAm) is also independent for each country, as is its variance.

3.3.3 NAC PFA Forecast model

WGNAS (ICES, 2009; 2012a) developed forecasts of the pre-fishery abundance for the non-maturing 1SW salmon (PFA) using a Bayesian framework that incorporates the estimates of lagged spawners and works through the fisheries at sea to determine the corresponding returns of 2SW salmon, conditioned by fisheries removals and natural mortality at sea. This model considers regionally-disaggregated lagged spawners and returns of 2SW salmon for the six regions of North America. The model is summarised in the Directed Acyclical Graph in Figure 3.3.1.1. The year is identified by the i index. The full code used for running the model is available on the WGNAS SharePoint site.

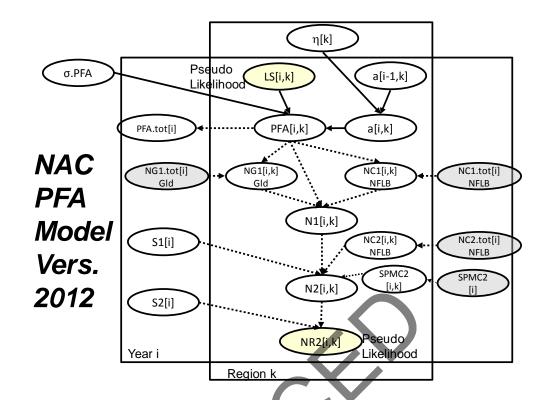


Figure 3.3.1.1. Directed Acyclical Graph (DAG) of the structure of the region disaggregated forecast model for 2SW salmon of North American origin. Ellipses in grey are observations (or pseudo-observations) derived from sampling programmes or from submodels (run-reconstruction).

Lagged spawners $LS_{i,k}$ represent the sum of smolt age adjusted annual spawners by region (k) that would be expected to contribute to the recruitment at sea prior to the fisheries (PFA) for year i, $LS_{i,k}$ are not directly observed but are estimated from the run-reconstruction submodel used to estimates returns and spawners to each of the six regions.

The probability distributions of LS (and returns of 2SW) by region are used as likelihood functions expressing comparative degrees of belief given the data and a probability model not explicitly specified in the current model. The probability distributions were drawn from the Monte Carlo simulations and assumed to be normal with known mean (LS.m) and precision (1/variance) (tau.LS). The use of this distribution as a likelihood function is equivalent to assuming a pseudo-observation equal to LS.m issuing from a sampling distribution with mean and precision equal to LS and tau.LS (Michielsens *et al.*, 2008).

$$LS.m_{i,k} \sim N \; (LS_{i,k}, \; tau.LS_{i,k})$$

The LS.m_i,k (mean) and tau.LS_{i,k} (precision) were derived assuming the lagged spawner values issued from a normal distribution characterized by the 95% confidence interval range statistics retained from the Monte Carlo simulations of returns.

Similarly, the returns of 2SW salmon to the six regions (NR2 $_{i,k}$) are not directly observed, but estimated from the run-reconstruction model. The probability distributions were assumed to be normal with known mean NR2.m and variance tau.NR2. As with the LS variable, the NR2 were treated as pseudo-observations equal

to NR2.m issuing from normal sampling distributions with means and variances equal to NR2 and tau.NR2.

$$NR2.m_{i,k} \sim N (NR2_{i,k}, tau.NR2_{i,k})$$

In between the lagged spawners and returns as 2SW salmon, the catches in the various sea fisheries and conditioning for natural mortality as the fish move from the time of the PFA to homewaters, are incorporated (Figure 3.3.1.1). The catches in the commercial fisheries of West Greenland and the Newfoundland and Labrador commercial and coastal fisheries (NG1.tot, NC1.tot and NC2.tot) are not directly observed, but estimated with error. The catches are converted to numbers of fish of 1SW non-maturing and 2SW fish based on the characteristics of the fish in the catch. Their (prior) probability distributions are obtained from catch statistics according to a formal structure included in the model.

Catches of large salmon (assumed to be 2SW salmon) from the St Pierre & Miquelon fisheries (SPMC) are also included in the model as point estimates.

The natural mortality in the post-PFA time point was assumed constant between years, centred on an instantaneous rate value of 0.03 per month (95% confidence interval range of 0.02 to 0.04).

For the NAC 2SW component, the model is fitted to an historical dataseries of lagged eggs starting from 1978. Although the return and spawner estimates for NAC begin in 1971, the lagged eggs are only available from 1978 due to the smolt age distributions (one to six years).

The years are modelled independently conditionally on the lagged spawners and yearly productivity parameters. The lagged spawners to PFA ratios (productivity) are modelled dynamically, i.e. assuming they are sequentially dependent within a region and attempts to take into account the most significant sources of uncertainty. The DAG for the model is shown in Figure 2.

PFA is assumed to be proportional to lagged-spawners (LS), with i.i.d. lognormal errors, and is modelled separately for each region (k = 6). The first year in the timeseries (t) is 1978 for lagged spawners (due to the range of smolt ages 1 to 6 for NAC and the start of the spawner time-series in 1970) and the last year of lagged spawner data is for the 2017 PFA year. The PFA can be modelled for 1978 to 2013 (the last PFA year for which returns of 2SW salmon have been estimated back to rivers in 2014).

$$\begin{aligned} & PFA_{i,k} = LogN(\overline{PFA_{i,k}}, \sigma.PFA^2) \\ & \overline{PFA_{i,k}} = \log(LS_{i,k}) + a_{i,k} + \varepsilon_{i,k} \\ & \varepsilon_{i,k} \sim N(0, \sigma.PFA^2) \end{aligned}$$

The total PFA is calculated as the sum of the regional PFA's (k = 6). The proportion of the total PFA in each region is calculated directly as:

$$p.PFA_{i,k} = PFA_{i,k} / PFA.tot_i$$

A non-informative prior is assumed for $\sigma.PFA^2$ (1/ $\sigma.PFA^2 \sim gamma(0.01, 0.01)$

The proportionality coefficient (log) $a_{i,k}$ between LS_{i,k} and PFA_{i,k} for each region is modelled dynamically as a random walk with a year and region residual variation (

 $\eta_{i,k}$) assumed multivariate normal (MVN). The variance covariance matrix (Σ) allows for correlations among regional productivity values reflecting that the fish share a common marine environment during part of their life cycle and that there are regional specificities in the evolution of the freshwater or the marine coastal environment.

$$a_{i+1,k} = a_{i,k} + \eta_{i,k}$$
$$\eta_{i,k} \stackrel{iid}{\sim} MVN(0, \Sigma)$$

The common yearly evolution of a_i is the mean of annual a across regions:

$$a.y_i \leftarrow mean(a_{i,k})$$

This parameterization of the covariance of the proportionality coefficient differs from an earlier version of the model for which the proportionality (log) coefficient $a_{i,k}$ between LS_{i,k} and PFA_{i,k} for each region dynamically as a random walk model with the addition of a regionally common annually varying parameter (e.y_i).

$$a_{i+1,k} = a_{i,k} + e.y_{i+1} + \omega_{i,k} \quad with \quad \omega_{i,k} \stackrel{i.i.d}{\sim} N(0, a.\sigma_k^2)$$

$$e.y_i \stackrel{i.i.d}{\sim} N(0, y.\sigma^2)$$

The correlation matrix of a among the regions is calculated from the covariance matrix:

- 1) the precision matrix is inverted to produce the covariance matrix;
- 2) the covariance matrix is transformed to the correlation matrix.

The positive-definite matrix (T, the precision matrix) is inverted:

```
covariance matrix <- solve(T)
correlation matrix <- cov2cor(b)</pre>
```

The dynamic component of the model requires initialization for the first year (i = 1978) and an uninformative prior is assumed:

$$a_{1,k} \stackrel{i.i.d}{\sim} N(0, 100)$$

The models are fitted and forecasts derived in a consistent Bayesian framework under the OpenBUGS 3.0.3 software (http://mathstat.helsinki.fi/openbugs/; Lunn et al., 2000).

3.3.4 Summary of NAC and NEAC forecast models

The data inputs and models currently used by WGNAS for forecasting and provision of catch advice differ between the Commission areas; outline details are summarised in the text table below.

	NAC	NEAC
Data inputs		
Time period of data	1978 on	1978 on for southern NEAC 1991 on for northern NEAC
Spatial aggregation	Separately for six regions of North America	By southern and northern stock complexes & NEAC countries
Age components	2SW salmon component only	1SW and MSW age components
Spawners	Lagged spawners by region for 2SW salmon only	Lagged eggs by sea age component for the southern and northern complexes/ country
Returns	Returns by region of 2SW salmon only	Returns of 1SW and MSW age components by stock complex / country
Model structure		
Spatial aggregation	Spawners and returns of 2SW salmon for six regions	Spawners and returns for two sea age components for the southern and northern NEAC complexes / countries
Dynamic function	Random walk dynamic	Random walk dynamic
	Region-specific recruitment rates linked with an annual recruitment rate variable	Sea age specific recruitment rates linked with a probability of maturing variable
Latent variables of interest	PFA 1SW non-maturing Recruitment rate by region and year	PFA 1SW maturing and PFA 1SW non-maturing by stock complex/country Recruitment rate by sea age component and the probability of maturing variable
Forecast years	Four years	Five years; i.e. the present year -1, the present year, and the next three years (y-1 is a forecast, as the MSW stock

3.4 The development of a risk analysis framework for catch advice

3.4.1 Introduction

The provision of catch advice in a risk framework involves incorporating the uncertainty in all the factors used to develop the catch options (ICES, 2002). The ranges in the uncertainties of all the factors will result in assessments of differing levels of precision. The analysis of risk involves four steps:

- 1) identifying the sources of uncertainty;
- 2) describing the precision or imprecision of the assessment;
- 3) defining a management strategy; and
- 4) evaluating the probability of an event (either desirable or undesirable) resulting from the fishery action.

The uncertainties have been identified and quantified in the assessment of PFA for salmon stocks in both the NAC and NEAC areas. NASCO's strategy for the management of salmon fisheries is based upon the principle of ensuring that stocks are above CLs (defined in terms of spawner escapement or egg deposition) with a high probability. The undesirable event to be avoided is that the spawning escapement after the fisheries will be below the CLs.

3.4.2 Catch advice and risk analysis framework for the West Greenland fishery

A risk framework for the provision of catch advice for the West Greenland fishery has been applied since 2003 (ICES, 2003) and has been subject to a number of subsequent updates. The current procedure is outlined below. This involves estimating the uncertainty in meeting defined management objectives at different levels of catch (catch options). The risk framework has been formally accepted by NASCO.

Two stock complexes are of relevance to the management of the West Greenland fishery; non-maturing 1SW fish from North America and non-maturing 1SW fish from southern NEAC. The risk assessments for the two stock complexes are developed in parallel and then combined at the end of the process into a single summary plot or catch options table. The risk analysis proceeds as illustrated in the flowchart in Figure 3.4.2.1).

The primary inputs to the risk analysis for the complex at West Greenland are:

- PFA forecast for the year of the fishery; PFANA and PFANEAC;
- Harvest level being considered (t of salmon);
- Conservation spawning limits or alternate management objectives; and
- The post-fishery returns to each region.

The risk analysis of catch options incorporates the following input parameter uncertainties: (i) the uncertainty of the pre-fishery abundance forecast, (ii) the uncertainty in the biological parameters used to translate catches (weight) into numbers of salmon, and (iii) the uncertainty in attaining the conservation requirements simultaneously in different regions.

The uncertainty in the PFA_{NA} and PFA_{NEAC} is accounted for in the forecast approach described above. The number of 1SW non-maturing fish of North American and European origin in a given catch (t) is conditioned by the continent of origin of the fish (prop_{NA}, prop_E), by the average weight of the fish in the fishery (Wtallages), and by the proportion 1SW non-maturing fish in the respective continent of origin catches. These parameters define how many fish originating in North America and Europe are expected in the fishery harvests. For a level of fishery under consideration, the weight of the catch is converted to number of fish of each continent of origin using the following equation:

$$C1SW_{c} = \frac{t \ X \ propC}{ACF * \ (propNA \ X \ Wt1SW_{NA} + propE \ X \ Wt1SW_{E})}$$

where: $C1SW_C$ is the catch (number of fish) of 1SW salmon originating in continent C (either North America or Europe), t is the fishery harvest at West Greenland in kg, propC is the proportion of the 1SW salmon harvest which originates from continent C, $Wt1SW_{NA}$ and $Wt1SW_E$ are the average weight (kg) in the fishery of a 1SW salmon of

North American and European origin, respectively, and *ACF* is the age correction factor by weight for salmon in the fishery which are not at age 1SW.

Since these parameters are not known for the forecast years of interest, they are estimated from previous values. Thus, propna (and propneac as 1 - propna) are drawn randomly from observed values of the past five years taking account of uncertainty due to sample sizes. For the other parameters, it is assumed that the parameters for Wtallages and the proportion non-maturing 1SW in the catch by continent of origin could vary uniformly within the values observed in the past five years.

For a level of fishery under consideration, the weight of the catch is converted to fish of each continent of origin and subtracted from one of the simulated forecast values of PFA_{NA} and PFA_{NEAC}. The fish that escape the Greenland fishery are immediately discounted by the fixed sharing fraction (F_{na}) historically used in the negotiations of the West Greenland fishery. The sharing fraction chosen is the 40%:60% West Greenland:North America split. The same sharing arrangement has been used for NEAC stocks (ICES, 2003). [Any sharing fraction could be considered and incorporated at this stage of the risk assessment].

After the fishery, fish returning to homewaters are discounted for natural mortality from the time they leave West Greenland to the time they return to rivers. For North America this is a total of eleven months at a rate of M = 0.03 (equates to 28.1% mortality). For southern NEAC stocks this is a total of eight months at a rate of M = 0.03 (equates to 21% mortality). The fish that survive to North American homewaters are then distributed among the regions based on the regional proportions of lagged spawners for the last five years when estimates of spawners were available. The uncertainty in the regional proportions is characterised by drawing at random from a uniform distribution defined by the minimum and maximum regional ranges from the five years and calculating the average proportion for each of the six regions in North America.

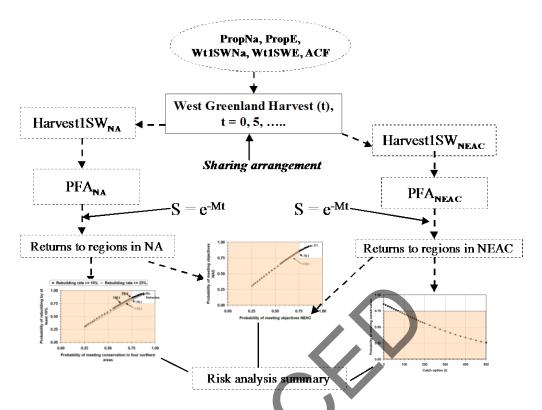


Figure 3.4.2.1. Flow chart summarising risk analysis for catch options at West Greenland using the PFANA and the PFANEAC predictions for the year of the fishery. Inputs with solid borders are considered known without error. Estimated inputs with observation error that is incorporated in the analysis have dashed borders. Solid arrows are functions that introduce or transfer without error whereas dashed arrows transfer errors through the components.

The final step in the risk analysis of the catch options involves combining the conservation requirement or alternate management objectives with the probability distribution of the returns to North America for different catch options. Estimated 2SW returns to each region are compared to the conservation objectives of Labrador, Newfoundland, Québec, and Gulf. Estimated returns for Scotia-Fundy are compared to the objective of achieving an increase of 25% relative to average returns of the base period, 1992–1996. For the USA, the management objective was revised in 2014 (ICES, 2014). Estimated returns for the USA are now compared to the objective of achieving 2SW adult returns of 4549 or greater. The advice to fisheries managers is presented as a probability plot (or table) of meeting or exceeding the objectives relative to increasing harvest levels at West Greenland.

ICES has adopted, a risk level of 75% of simultaneous attainment of seven management objectives (ICES, 2003) as part of an agreed management plan for the West Greenland fishery. The same level of risk aversion is applied for catch advice for homewater fisheries on the North American stock complex.

The catch advice for the West Greenland fishery is currently tabulated to show the probability of each management unit achieving its CL (or alternative reference level) individually and the probability of this being achieved by all management units simultaneously (i.e. in the same given year) (e.g. ICES, 2012a). This allows managers to evaluate both individual and simultaneous attainment levels in making their management decisions. Table 3.4.2.1 provides an example of catch options for West Greenland for the years 2012 to 2014 (ICES, 2012a). An updated catch options table

for 2015 to 2017 is provided in Section 5.3 of the 2015 Working Group report (see above).

The models currently used by WGNAS in developing catch advice are considered to provide consistent characterisation of the status and expectations for Atlantic salmon in the North Atlantic. Compared to previous models used by WGNAS prior to 2009, the Bayesian models provide more flexibility, are consistent with the emerging emphasis on such approaches in natural resource assessment, and can provide management advice consistent with the probability of achieving management objectives.

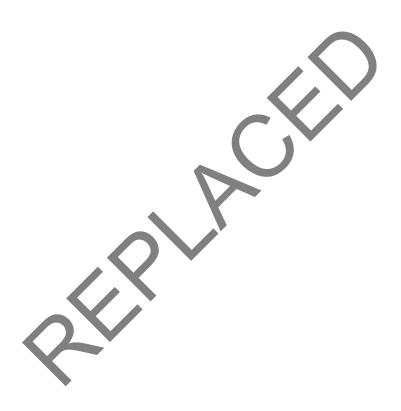


Table 3.4.2.1. Catch options tables for mixed-stock fishery at West Greenland by year of PFA, 2012 to 2014.

2012 Сатсн	PROBABIL	ITY OF MEETI	NG OR EXCE	EDING REGION	N-SPECIFIC	MANAGEMEN	T OBJECTIVES	
OPTION (T)	LAB	NFLD	QC	GULF	SF	USA	S-NEAC	ALL
0	0.45	0.86	0.71	0.50	0.15	0.89	0.98	0.06
10	0.42	0.84	0.67	0.48	0.14	0.88	0.98	0.05
20	0.40	0.83	0.63	0.45	0.13	0.87	0.98	0.05
30	0.38	0.81	0.59	0.42	0.12	0.85	0.98	0.04
40	0.36	0.78	0.54	0.40	0.12	0.83	0.98	0.04
50	0.34	0.76	0.50	0.38	0.11	0.81	0.98	0.03
60	0.32	0.73	0.46	0.36	0.10	0.79	0.98	0.03
70	0.30	0.70	0.42	0.33	0.09	0.77	0.98	0.03
80	0.28	0.67	0.39	0.31	0.08	0.74	0.98	0.03
90	0.26	0.64	0.35	0.29	0.08	0.72	0.97	0.02
100	0.24	0.60	0.32	0.27	0.07	0.68	0.97	0.02
2013 Сатсн	PROBABIL	ITY OF MEETI	NG OR EXCE	EDING REGIO	-SPECIFIC	MANAGEMEN	T OBJECTIVES	
OPTION (T)	LAB	NFLD	QC	GULF	SF	USA	S-NEAC	ALL
0	0.48	0.78	0.73	0.50	0.25	0.75	0.95	0.08
10	0.46	0.76	0.70	0.48 0.24		0.73	0.95	0.07
20	0.44	0.75	0.67	0.46 0.23		0.72	0.95	0.06
30	0.42	0.73	0.63	0.44 0.22		0.70	0.95	0.06
40	0.41	0.70	0.60	0.42	0.21	0.68	0.95	0.06
50	0.39	0.68	0.56	0.40	0.20	0.66	0.94	0.05
60	0.37	0.65	0.53	0.38	0.19	0.64	0.94	0.05
70	0.35	0.63	0.50	0.36	0.18	0.62	0.94	0.05
80	0.33	0.60	0.47	0.34	0.17	0.59	0.94	0.04
90	0.31	0.57	0.44	0.32	0.16	0.57	0.94	0.04
100	0.30	0.54	0.41	0.31	0.15	0.55	0.94	0.04
2014 Сатсн	PROBABIL	ITY OF MEETI	NG OR EXCE	EDING REGION	N-SPECIFIC	MANAGEMEN	T OBJECTIVES	
OPTION (T)	LAB	NFLD	QC	GULF	SF	USA	S-NEAC	ALL
0	0.56	0.78	0.75	0.55	0.20	0.86	0.94	0.08
10	0.55	0.77	0.73	0.53	0.20	0.85	0.94	0.08
20	0.53	0.75	0.70	0.51	0.19	0.84	0.94	0.07
30	0.52	0.73	0.67	0.49	0.18	0.83	0.94	0.07
40	0.50	0.71	0.64	0.47	0.17	0.82	0.94	0.06
50	0.48	0.69	0.62	0.46	0.17	0.81	0.94	0.06
60	0.46	0.67	0.59	0.44	0.16	0.79	0.94	0.06
70	0.45	0.65	0.56	0.42	0.16	0.77	0.94	0.05
80	0.43	0.63	0.54	0.41	0.15	0.76	0.94	0.05
90	0.42	0.61	0.51	0.39	0.14	0.74	0.94	0.05

3.4.3 Catch advice and risk analysis framework for the Faroes fishery

3.4.3.1 Outline of the risk framework and management decisions required

There is currently no agreed framework for the provision of catch advice for the Faroes fishery adopted by NASCO. However, NASCO has asked ICES, for a number of years, to provide catch options or alternative management advice with an assessment of risks relative to the objective of exceeding stock conservation limits for salmon in the NEAC area. An initial risk framework that could be used to provide and evaluate catch options for the Faroes fishery was outlined by WGNAS in 2010 (ICES, 2010a). This was based on the method currently used to provide catch advice for the West Greenland fishery, which involves estimating the uncertainty in meeting defined management objectives at different catch levels (TAC options). The Faroes risk framework was developed further at subsequent WGNAS meetings (ICES, 2011; 2012a) and the current proposed procedure is outlined below.

A number of decisions are required by managers before full catch advice could be provided (ICES, 2011; 2012a). Specifically, ICES has indicated that NASCO would need to agree upon the following issues before the risk framework could be finalised:

- season (January to December or October to May) to which any TAC should apply;
- share arrangement for the Faroes fishery;
- choice of management units for NEAC stocks; and
- · specification of management objectives.

In developing an indicative risk framework, WGNAS has made pragmatic choices regarding these issues:

Faroes fishing season: A decision is required on the period to which any TAC for the Faroes fishery would apply. The Faroes fishery has historically operated between October/November and May/June, but the historical TACs applied to a calendar year. This means that two different cohorts of salmon of each age class (e.g. two cohorts of 1SW salmon, etc.) were exploited under each TAC. ICES (2011) recommended that NASCO manage any fishery on the basis of fishing seasons operating from October to June, and catch advice should be provided on this basis.

Sharing agreement: The 'sharing agreement' establishes the proportion of any harvestable surplus within the NEAC area that could be made available to the Faroes fishery through the TAC. In effect, for any TAC option being evaluated for the Faroes, it is assumed that the total harvest would be the TAC divided by the Faroes share. WGNAS has proposed using a share allocation derived using the same approach and baseline period (1986–1990) as for West Greenland (ICES, 2010a). This gave a potential share allocation of 7.5% to Faroes. Following discussion within NASCO, one Party proposed an alternative baseline period of 1984–1988, which would give a share allocation of 8.4% to Faroes. In the absence of further advice from NASCO, WGNAS has applied a value of 8.4%.

Choice of management units: ICES (2010a) noted that the stock complexes currently used for the provision of NEAC catch advice (southern NEAC and northern NEAC) are significantly larger than each of the six management units used for North American salmon (2SW only) in the catch advice for the West Greenland fishery. Basing an assessment of stock status on these large units greatly increases the risks to

individual NEAC river stocks or groups of stocks that are already in a more depleted state than the average.

For the provision of catch advice on the West Greenland fishery, the total CL for NAC (2SW salmon only) of about 152 000 fish is assessed in six management units, which means that each unit has an average CL of about 25 000 salmon. In contrast, the total CLs for each of the NEAC stock complexes are:

Northern NEAC 1SW-	158 223
Northern NEAC MSW-	131 356
Southern NEAC 1SW-	565 183
Southern NEAC MSW-	275 549

The NEAC stock complexes are therefore between eight and 25 times the size of the average NAC ones. There is also wide variation in the size and status of stocks both within and among the NEAC national stock groups. WGNAS recommended (ICES, 2012a) that the NEAC catch advice should be based on more management units than are used at present, but also noted that there are practical limitations on the extent to which the assessments can be disaggregated, since the availability of information on the composition of the catch at Faroes constrained the selection of management units. In 2013, WGNAS (ICES, 2013) proposed a method to estimate the stock composition of the Faroes catch at a national level based on tag returns and the PFA estimates, but did not consider it appropriate to extend this to stock complexes smaller than this. Genetic stock assignment studies are underway to analyse scale samples collected at Faroes, but these are not expected to facilitate disaggregation below this level. In addition, other parameter values used in the assessment are currently only available for the total fishery and not smaller stock complexes.

In providing indicative catch advice with the new framework, WGNAS considered that it would be informative to managers to provide catch options tables for the ten NEAC countries as well as for the four stock complexes and has therefore run the risk framework using management units based on countries.

Management objectives: The management objectives provide the basis for determining the risks to stocks in each management unit associated with different catch options. The NASCO agreement on the adoption of a Precautionary Approach (NASCO, 1998) calls for the 'formulation of pre-agreed management actions in the form of procedures to be applied over a range of stock conditions', indicating that the management objectives (e.g. the required probability of exceeding the CL) should be agreed in advance of specific management proposals being considered.

At the request of NASCO, WGNAS considered the implications of applying probabilities of achieving CLs to separate management units vs. the use of simultaneous probabilities; this issue was outlined in detail in ICES (2013).

The probability of simultaneous attainment of management objectives in a number of separate management units is roughly equal to the product of the probabilities of individual attainment for each management unit. The probability threshold for each individual management unit might reasonably be set at a fixed level unless there are specific reasons for adopting an alternative (e.g. for stock rebuilding). ICES (2012) recommended that an appropriate probability level for individual stock complexes would be 95%. This individual probability level can be applied to each management

unit regardless of the number of units used; however, this is less obvious for the probability of simultaneous attainment, as explained next.

Management decisions for the West Greenland fishery have been based on a 75% probability of simultaneous attainment of CLs. For a given probability of achieving individual stock CLs, the probability of simultaneous attainment decreases rapidly as the number of management units considered increases. For the example of 20 management units (e.g. two age groups from each of ten countries), the use of the simultaneous probability level applied for West Greenland (75%) would correspond to the probability of individual stocks meeting the CLs being 98.6% or higher, assuming the same individual probability for all stocks. The use of a 95% probability level for meeting the CLs individually in the 20 management unit example, implies a simultaneous attainment probability of about 36%, i.e. there would be a 64% chance that at least one stock failed to meet its CL in any given year. On the other hand, the use of a 75% probability of simultaneous attainment could result in a fishery being advised when the individual probability of one management unit is as low as 75% if all the other management units have a 100% chance of meeting the CL (as in that case, the probability of simultaneous attainment would still be 75%). This may not be an acceptable risk for managing multiple river stocks.

WGNAS considered that the probability of simultaneous attainment can provide useful information to managers of the risk of failing to meet CLs in at least one stock in the MSF. However, as the management units being considered by NASCO for managing the MSF at Faroes are still very large and each unit encompasses a large number of individual river stocks, choosing a high probability level (such as 95%) of attaining CLs in individual units would be less risky to individual stocks than the use of a simultaneous attainment objective set at the value used for the West Greenland fishery.

On the basis of these considerations, WGNAS provided both individual probabilities and the probability of simultaneous attainment of the management units in the catch options tables (ICES, 2013). ICES recommends that management decisions should be based principally on a 95% probability of attainment of CLs in each stock complex individually. The simultaneous attainment probability may also be used as a guide, but managers should be aware that this will generally be quite low when large numbers of management units are used (as illustrated above, in the example with 20 management units).

3.4.3.2 Modelling approach for the catch options risk framework

The basic model for assessing each catch option within the risk framework is the same for both stock complexes and at a country level (ICES, 2013). The PFA forecasts derived in the Winbugs model are transferred to the risk framework model run in 'R'. The estimates and distributions of the PFA estimates used in the risk framework are derived by taking the first 50 000 values from the Winbugs posterior forecast simulations. Parameters in the following description that are marked with an '*' in the equations have uncertainty around them generated by means of 50 000 random draws from the annual values observed from the sampling programmes conducted in the Faroes between the 1983/1984 and 1990/1991 fishing seasons. They therefore contribute to the estimation of the probability density function around the potential total harvest arising from each TAC option. When the assessment is run at a national level, the number of draws has to be limited to 25 000 because of memory limitations in 'R'.

The modelling procedure involves:

• estimating the total number of 1SW and MSW salmon that could be killed as a result of any TAC at Faroes, including catches in homewaters;

- adjusting these to their equivalent numbers at the time of recruitment to the Faroes fishery;
- subtracting these from the PFA estimates for maturing and non-maturing 1SW salmon in the appropriate years;
- assessing the results against the SERs (i.e. the CLs adjusted to the time of recruitment to the Faroes fishery).

The TAC option (T) is first divided by the mean weight (Wt*) of salmon caught in the Faroes fishery to give the number of fish that would be caught, and this value is converted to numbers of wild fish (Nw) by multiplying by one minus the proportion of fish-farm escapees in samples taken from the Faroes catch (pE*) observed in historical sampling programmes. A correction factor (C = 0.63) is applied to the proportion of fish-farm escapees to take account of reductions in the numbers of farm escapees over the past 20 years based on observations in Norwegian coastal waters:

$$Nw = T / Wt^* x (1 - (pE^* x C))$$

This value is split into numbers by sea age classes (1SW and MSW) according to the proportion of each age group (pAi*, where 'i' is 1SW or MSW) observed in historical catch sampling programmes at Faroes. In the past, there has also been a requirement to discard any fish less than 60 cm total length caught in the Faroes fishery, and 80% of these fish were estimated to die, so these mortalities are also added to the 1SW catch. Thus:

```
Nw1SW = Nwtotal \ x \ pA1SW^* + (Nwtotal \ x \ pD^* / (1 - pD^*) \ x \ 0.8) and
```

 $NwMSW = Nwtotal \times pAMSW^*$

where: pD* is the proportion of the total catch that is discarded (i.e. fish <60 cm total length).

Further corrections are made to the 1SW and MSW numbers to reduce the 1SW total to take account of the proportion that will not mature as 1SW fish and to add the survivors from this group to the MSW fish in the following year. For the first catch advice year the number added to the MSW total is adjusted to the TAC of the current season (i.e. zero in 2012/2013). Thus:

```
Nw1SW = Nw1SW \times pK * and NwMSW = NwMSW + Nw1SW \times (1-pK*)
```

where 'pK'' is the proportion of 1SW salmon that are expected to mature in the same year (0.78) derived from experimental studies conducted in the 1980s (Youngson and Webb, 1993).

The numbers in each age group are then divided among the management units by multiplying by the appropriate proportions (pUij), where 'i' denotes the age groups

and 'j' denotes the management units, and each of these values is raised by the Faroes share allocation (S) to give the total potential harvest (Hij) of fish from each management unit and sea age group:

$$Nwij = (Nwi \times pUij) / S$$

Finally, these values are adjusted for natural mortality so that they can be compared with the PFA forecasts and SER values from the mid-date of the fishery to the recruitment date by using an instantaneous monthly rate of mortality of 0.03.

These harvests are then subtracted from the stock forecasts (PFAij) for the management units and sea age groups and compared with the Spawner Escapement Reserves (SERij) to evaluate attainment of the management objective. In practice, the attainment of the management objective is assessed by determining the probability that PFAij – Hij – SERij is greater than zero. The SER is the number of fish that need to be alive at the time of the Faroes fishery to meet the CL when the fish return to homewaters; this equals the CL raised by the mortality over the intervening time. CLs and SERs are currently estimated without uncertainty.

3.4.3.3 Input data for the risk framework

The analysis estimates probability of each management unit achieving its SER (the overall abundance objective) for different catch options in the Faroes fishery (from 0 to 200 t). The analysis assumes:

- no fishery operated in the 2012/2013 season;
- the TAC allocated to Faroes is the same in each year and is taken in full;
- homewater fisheries also take their catch allocation in full.

The analysis requires the following input data for the catch that would occur at the Faroes if a TAC was allocated (full details are provided in ICES, 2013):

- mean weights;
- proportion by sea age;
- discard rates (fish less than 60 cm total length);
- proportion of fish-farm escapees;
- composition of catches by management unit;
- proportion of 1SW fish not maturing.

3.4.3.4 Indicative catch advice

Table 3.4.3.4.1 provides an example of catch options for the Faroes fishery for the seasons 2013/2014 to 2015/2016 (ICES, 2013). Equivalent tables were provided for both 1SW and MSW salmon for all NEAC countries, and WGNAS also estimates the exploitation rates that these TAC options would impose on each stock complex or national stock (ICES, 2013). Updated catch options tables for the seasons 2015/2016 to 2017/2018 are provided in Section 3.5 of the 2015 Working Group report (see above).

Table 3.4.3.4.1. Probability of northern and southern NEAC - 1SW and MSW stock complexes achieving their SERs independently and simultaneously for different catch options for the Faroes fishery in the 2013/2014 to 2015/2016 fishing seasons.

Catch options	TAC option	NEAC-N-	NEAC-N-	NEAC-S-	NEAC-S-	All
for 2013/14	(t)	1SW	MSW	1SW	MSW	complexes
season:	0	96.2%	99.8%	74.3%	75.6%	56.8%
	20	96.2%	99.2%	74.2%	69.8%	52.7%
	40	96.2%	98.2%	74.2%	63.9%	48.2%
	60	96.1%	96.3%	74.1%	57.9%	43.3%
	80	96.1%	93.4%	74.1%	52.1%	38.1%
	100	96.1%	89.3%	74.0%	46.6%	32.9%
	120	96.0%	84.3%	74.0%	41.7%	28.1%
	140	96.0%	78.4%	73.9%	36.8%	23.4%
	160	95.9%	71.6%	73.9%	32.5%	19.2%
	180	95.9%	64.6%	73.8%	28.5%	15.4%
	200	95.8%	57.6%	73.8%	25.0%	12.2%
Catch options	TAC option	NEAC-N-	NEAC-N-	NEAC-S-	NEAC-S-	All
for 2014/15	(t)	1SW	MSW	1SW	MSW	complexes
season:	0	94.6%	99.2%	75.4%	79.6%	59.0%
	20	94.6%	98.2%	75.3%	75.3%	55.8%
	40	94.6%	96.6%	75.3%	70.8%	52.0%
	60	94.5%	94.2%	75.2%	66.4%	48.0%
	80	94.4%	90.9%	75.2%	61.8%	43.6%
	100	94.4%	86.8%	75.1%	57.3%	38.9%
	120	94.3%	82.1%	75.1%	53.1%	34.4%
	140	94.3%	76.8%	75.0%	49.0%	30.1%
	160	94.3%	71.2%	75.0%	45.0%	25.9%
	180	94.2%	65.5%	74.9%	41.5%	22.1%
	200	94.2%	59.6%	74.9%	38.0%	18.6%
Catch options	TAC option	NEAC-N-	NEAC-N-	NEAC-S-	NEAC-S-	All
for 2015/16	(t)	1SW	MSW	1SW	MSW	complexes
season:	0	94.6%	98.5%	70.1%	79.7%	55.2%
	20	94.6%	97.2%	70.1%	76.0%	52.4%
	40	94.5%	95.1%	70.0%	72.2%	49.2%
	60	94.5%	92.3%	70.0%	68.4%	45.6%
	80	94.5%	89.0%	69.9%	64.6%	41.9%
	100	94.4%	85.0%	69.9%	60.7%	38.0%
	120	94.4%	80.6%	69.8%	57.1%	34.2%
	140	94.3%	75.7%	69.8%	53.5%	30.4%
	160	94.3%	70.6%	69.7%	50.0%	26.7%
		94.2%	65.4%	69.7%	46.8%	23.4%
	180					

3.5 Development of indicator frameworks to identify significant changes in previously provided multiannual management advice

3.5.1 Background

In support of the multiannual management advice that is provided for all three NASCO Commission Areas, NASCO asked ICES to provide an assessment of the minimal information needed to signal an unforeseen change in productivity for

stocks contributing to fisheries within each Commission area. A particular concern was that an increase in productivity may alter the reliability of the previously provided multiyear catch options and could result in unrealised harvest within various mixed-stock fisheries. Initial progress on this issue was presented to WGNAS in 2006 (ICES, 2006) and further developments were made by the Study Group on Establishing a Framework of Indicators of Salmon Stock Abundance [SGEFISSA] which met in November 2006 (ICES, 2007b) and reported to WGNAS in 2007 (ICES, 2007a). This resulted in the development of a suggested framework (Framework of Indicators - FWI) which could be used to indicate if any significant change in the status of stocks had occurred and thus confirm whether the previously provided multi-annual management advice was still appropriate.

The initial FWI was developed with both the Greenland and Faroes fisheries in mind, although the methodology only proved suitable for the West Greenland fishery and an alternative approach was subsequently developed for the NEAC area (ICES, 2011; 2012a; 2013). Thus, FWIs are now routinely applied in the interim (non-assessment) years of multiyear agreements for both NAC and NEAC to facilitate the management of the West Greenland and Faroes fisheries respectively. Both operate according to the timeline outlined in Figure 3.5.1.1. Outline descriptions of the two different schemes are provided below.

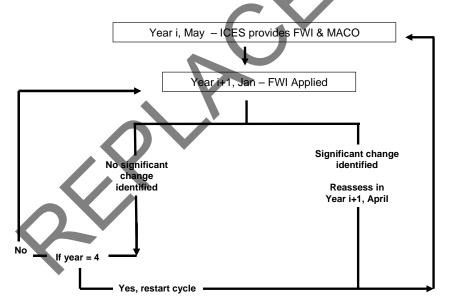


Figure 3.5.1.1. Timeline for employment of the Framework of Indicators (FWI). In Year i, ICES provides an updated FWI which re-evaluates the updated datasets and is summarized in an Excel worksheet. In January of Year i+1 the FWI is applied and two options are available depending on the results. If no significant change is detected, no re-assessment is necessary and the cycle continues to Year i+2. If no significant change is detected in Year i+2, the cycle continues to Year i+3. If a significant change is detected in any year, then reassessment is recommended. In that case, ICES would provide an updated FWI the following May. ICES would also provide an updated FWI if year equals 4. [MACO = multi-annual catch options].

3.5.2 Framework of Indicators (FWI) for the West Greenland Fishery

The process for developing and applying the FWI for the Greenland fishery consists of six general steps:

• <u>Definition of a significant change</u> - Define measurable criteria for what the statement "a significant change in the previously provided multi-annual management advice" represents.

- Evaluating historical relationships between indicators and variables of interest - Define and evaluate the historical relationships between numerous indicators and the variable of interest for individual rivers across all stock complexes.
- <u>Establishing threshold values</u> Define the threshold level (i.e. variable of interest level) that will satisfy the management objectives for each stock complex.
- <u>Decision rule determinations</u> Define and apply a standardised approach
 for determining the appropriate decision rule value. The decision rule
 should provide a signal if the variable of interest will be greater than or
 less than the threshold level with high precision.
- Combining Indicators within the Framework Define and apply a standardised approach for combining indicator datasets within and across stock complexes for future comparison against contemporary indicator values.
- Applying the FWI Define and apply a standard approach to input contemporary indicator values into the FWI to determine if there is likely to be a significant change in the previously provided management advice.

Each of these is considered in turn; full details are available in ICES (2007b).

3.5.2.1 Definition of a significant change

A significant change in the previously provided multiannual management advice is regarded as an unforeseen change in stock status that would alter the previously provided advice based on analysis of current population data obtained from various monitored populations across the North Atlantic. This would be indicated by a situation where stock abundance has increased to a level where a fishery could be recommended when no catch had previously been advised, or a decrease in stock abundance when catch options had been chosen.

For the fishery at West Greenland, ICES would recommend that a harvestable surplus exists within the West Greenland stock complex if there was a high probability (75%) that the following three objectives could be met simultaneously:

The conservation limits of the four northern regions of North America (Labrador, Newfoundland, Québec, and Gulf) were achieved.

There was a 25% increase in returns to the Scotia-Fundy region relative to the mean returns for the 1992–1996 period. For the USA, the management objective was revised in 2014 to correspond to recover objectives defined in the recovery plan for endangered Atlantic salmon stocks in the USA (ICES, 2014), this now requires that estimated 2SW adult returns are 4549 or greater.

The conservation limit for the Southern NEAC MSW complex was achieved.

3.5.2.2 Evaluating historical relationships between indicators and variable of interest

A number of variables were considered for inclusion as indicators in the FWI, but only two were considered sufficiently informative to be carried forward into the framework: adult returns (returns, catch or estimated PFA) and return rates (i.e.

smolt survival rates, marine survival). These are available, by sea age class, for a number of monitored rivers throughout the North Atlantic and can be directly related to the management objectives for the fishery.

3.5.2.3 Establishing threshold values

In keeping with the 75% probability of meeting or exceeding the objectives for the West Greenland catch options (see above), the 25th percentile of the return estimates of the six areas in North America are compared to the corresponding 2SW conservation limits of the four northern areas of North America, to the 25% increase objective for the Scotia- Fundy area, and to management objective of achieving 4549 or greater 2SW adult returns for the USA. For the southern NEAC non-maturing component, the 25th percentile of the PFA estimate of the southern NEAC non-maturing complex is compared to the spawning escapement reserve (SER) for the southern NEAC non-maturing complex.

3.5.2.4 Decision rule determinations

The procedure for analysing the relationships between the indicators and the returns of 2SW salmon or the PFA estimates was originally suggested by ICES (2006). The individual river catches, returns or return rates are lagged to correspond to the same smolt cohort for the 2SW returns to North America or to the PFA estimates for NEAC complexes. Bivariate plots of each indicator dataset relative to the 2SW returns or the PFA estimates are prepared. Upper and lower halves are defined by the management objective value for the corresponding geographic area in North America or the NEAC stock complexes as outlined above. Estimates of returns of 2SW or PFA estimates in the upper half correspond to years when the returns or PFA exceed the management objectives. Points in the lower half correspond to years when the returns or PFA are less than the management objective.

Left and right halves are defined by a sliding rule along the indicator range. An objective function that maximises the number of correct assignments (true highs and true lows) is used to define the indicator decision rule. The objective function also minimises the number of incorrect assignments (false highs and false lows, Figure 3.5.2.4.1).

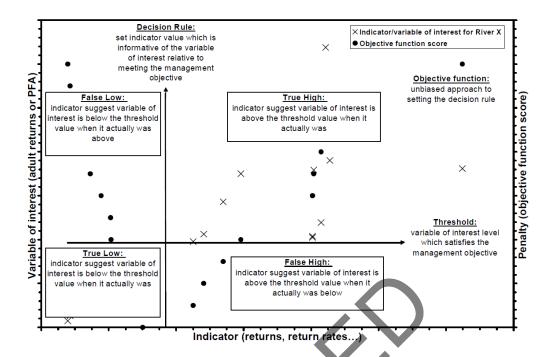


Figure 3.5.2.4.1. Example of Indicator/Variable of Interest exploratory graph identifying the threshold value, decision rule, penalty function and the four states (true high, true low, false high and false low).

The value of the indicator variable that minimises the sum of the penalty scores (i.e. minimises the number of incorrect assignments) is assigned as the decision rule for that dataset. Equal penalty weights are assigned to false highs (lower right quadrant) and false lows (upper left quadrant). Correct assignments are scored as zero. In the case when multiple minima occurred, the lowest indicator value among the low minima values is chosen.

Indicators are retained in the framework when they are evaluated as being informative of the magnitude of returns or PFA relative to the management objectives. These informative indicator datasets also have to meet the following two criteria to be retained:

- Expectation that the indicator variable would be available in future (in January), and
- A minimum of five observations are present in each of the correct quadrats (true low; true high).

3.5.2.5 Combining Indicators within the Framework

The probabilities of correct assignments are calculated for each of the true low and true high states for each of the indicator datasets retained. The respective probabilities correspond to the ratio of the correct assignment to all observations within the respective low and high indicator halves:

 $P(State_{low} \mid Indicator_{low})$ (i.e. true $low) = N(State_{low} \mid Indicator_{low}) / N$ $Indicator_{low}$

 $P(Statehigh \mid Indicatorhigh) \ (i.e. \ true \ high) = N(Statehigh \mid Indicatorhigh) \ / \ N$ Indicatorhigh

Indicator datasets are then pooled according to management objective/stock complex groupings. Each NAC stock complex (n=6) and the southern NEC non-maturing stock complex are pooled separately as these stock complexes relate to the management objectives for the West Greenland fishery.

3.5.2.6 Applying the FWI

To apply the FWI, the most recent year's indicator value for each of the retained indicator datasets is compared to the decision rule as determined from the historical datasets. If the contemporary indicator value is low relative to the decision rule, it is assigned a value of -1. If the value is high, it is assigned a score of +1. Multiple indicators within the stock complex groupings are then combined by arithmetic average of the product of the indicator value (-1, +1) and the probability of a correct assignment corresponding to the true low or true high states. An average geographic area or stock complex score equal to or greater than zero suggests there is a likelihood of meeting the management objective for that grouping based on the historic relation between the variable of interest (adults returns or PFA) and the indicators evaluated.

If the scores for all the groupings within a fishery complex are greater than zero, then there is a likelihood that all the management objectives for that fishery will be met. Under that scenario, the multiyear management advice should be reassessed. When the score(s) for one or more of the groupings is less than zero, there is unlikely to be a significant change in the management advice and there would be no need for a reassessment.

SGEFISSA (ICES, 2007b) developed a spreadsheet template FWI (see example at Figure 3.5.2.6.1) in which the underlying variable of interest/ indicator dataset relationships and decision rules are summarised and collated according to the specific management objectives for each fishery. This provides one of two conclusions for the user:

- 1) No significant change identified by the indicators;
- 2) Reassess.

	Catch Advice	Catch opt (Yes = 1, I			0					
				rall Reco						
		No S	ignificant	Change I	dentified	by Indica	ators			
		2011	Ratio Value to				Indicator	Probability of Correct	Indicator	Managemei Objective
Geographic Area USA	River/ Indicator Penobscot 2SW Returns	Value 2368		Threshold		True High 92%	State	Assignment	Score	Met?
JOA	Penobscot 1SW Returns	741	167% 197%	1415 377	100% 83%	88%	1 1	0.92 0.88	0.92 0.88	
	Penobscot 2SW Survival (%)	0.39	170%	0.23	100%	60%	1	0.6	0.6	
	Penobscot 1SW Survival (%)	0.12	133%	0.09	85%	73%	1	0.73	0.73	
	Narraguagus Returns	196	196%	100	95%	61%	1	0.61	0.61	
	possible range		.===		-0.93	0.75				.,
	Average		173%						0.75	Yes
Scotia-Fundy	Saint John Return Large	294	9%	3 329	96%	100%	-1	0.96	-0.96	
,	Lahave Return Large	146	51%	285	77%	85%	-1	0.77	-0.77	
	St. Mary's Return Large	14	6%	221	100%	73%	-1	1	-1	
	North Return Large	1 193	168%	712	95%	67%	1	0.67	0.67	
	Saint John Return 1SW	582	26%	2 276	86%	80%	-1	0.86	-0.86	
	LaHave Return 1SW	565	34%	1 679	94%	67%	-1	0.94	-0.94	
	St. Mary's Return 1SW	331	16% 59%	2 038 0.22	95% 95%	93% 81%	-1 -1	0.95 0.95	-0.95 -0.95	
	Saint John Survival 2SW (%) Lahave Survival 2SW (%)	0.13 0.88	367%	0.22	95% 81%	81%	-1 1	0.95	-0.95 0.81	
	Saint John Survival 1SW (%)	0.88	16%	0.24	86%	73%	-1	0.86	-0.86	
	Lahave Survival 1SW (%)	0.72	50%	1.44	92%	78%	-1	0.92	-0.92	
	Liscomb Survival 2SW (%)	0.03	60%	0.05	86%	91%	-1	0.86	-0.86	
	East Sheet Harbour Survival 2SW (%)	0.005	25%	0.02	67%	82%	-1	0.67	-0.67	
	possible range				-0.88	0.81				
	Average		68%				_		-0.64	No
Gulf	Miramichi Return 2SW	28 977	183%	15 800	100%	85%		0.85	0.85	
	Miramichi Return 1SW	45 880	110%	41 790	89%	67%	1	0.67	0.67	
	possible range				-0.95	0.76				
	Average		147%						0.76	Yes
	Occupation But and I come	0.045	4070/	0.000	2004	000/		2 2 2 2	0.00	
Quebec	Cascapédia Return Large	3 815	167%	2 280	69%	92% 81%	M	0.92	0.92	
	Bonaventure Return Large Grande Rivière Return Large	1 259 533	85% 121%	1 479 442	75% 100%	94%		0.75 0.94	-0.75 0.94	
	Saint-Jean Return Large	688	91%	758	86%	89%	△ -1	0.86	-0.86	
	Dartmouth Return Large	1 171	155%	756	86%	89%	1	0.89	0.89	
	Madeleine Return Large	996	153%	653	70%	93%	1	0.93	0.93	
	Sainte-Anne Return Large	871	201%	433	67%	88%	1	0.88	0.88	
	Godbout Return Large	694	108%	641	86%	100%	1	1	1	
	De la Trinite Return Large	317	82%	385	75%	100%	-1	0.75	-0.75	
	York Return Return Large	1 585	113%	1405	63%	83%	1	0.83	0.83	
	Grande Rivière Return Small	237	119%	199	59%	80%	1	0.8	0.8	
	Saint-Jean Return Small Godbout Return Small	343 623	87% 123%	394 508	53% 85%	80% 92%	-1 1	0.53 0.92	-0.53 0.92	
	De la Trinite Return Small	949	238%	399	89%	83%	1	0.83	0.83	
	De la Trinite Return Small De la Trinite Survival Large (%)	0.76	155%	0.49	88%	96%	1	0.96	0.96	
	De la Trinite Survival Small (%)	2.54	170%	1.49	63%	89%	i	0.89	0.89	
	Saint-Jean Survival Small (%)	1.86	258%	0.72	100%	64%	1	0.64	0.64	
	possible range				-0.77	0.88				
	Average		143%						0.50	Yes
Newfoundland	Exploits Return Small	34 085	137%	24 924	83%	56%	1	0.56	0.56	
Newtoundiand	Middle Brook Return Small	2 642	141%	1 868	84%	63%	1	0.63	0.63	
	Torrent Return Small	2 784	67%	4 154	94%	64%	-1	0.94	-0.94	
	possible range	2101	01 70		-0.87	0.61		0.01	0.01	
	Average		115%						0.08	Yes
		,								
_abrador	possible range									
	possible range Average								NA	Unknowr
									.47	CHAHOWI
Southern NEAC	<i>/</i>) \									
	possible range									
	Average								NA	Unknow

Figure 3.5.2.6.1. Framework of indicators spreadsheet for the West Greenland fishery. For illustrative purposes, the 2011 value of returns or survival rates for the 40 retained indicators is entered in the cells corresponding to the annual indicator variable values.

If no significant change has been identified by the indicators, then the multiyear catch advice for the year of interest could be retained. If a significant change is signalled by the indicators, the response is to reassess.

The framework spreadsheet is designed to capture both fishing and non-fishing scenarios:

- Multiyear advice provides no catch options greater than zero but indicators are suggesting that the management objectives may be met (conclusion: Reassess);
- Multiyear advice provides catch options greater than zero but the indicators suggest the management objectives may not be met (conclusion: Reassess).

There are two steps required by the user to run the framework. The first step in the framework evaluation is to enter the catch advice option (i.e. tonnes of catch) for the West Greenland fishery. This feature provides the two way evaluation of whether a change in management advice may be expected and a reassessment would be required. The second step is to enter the values for the indicator variables in the framework for the year of interest. The spreadsheet evaluation update is automated and the conclusion is shown in the row underneath "Overall Recommendation".

The conclusions from the framework evaluation are based on whether there is simultaneous achievement of the management objectives in the six stock areas of North America and the southern NEAC 1SW non-maturing complex. If there are no indicator variables for a geographic area, the attainment of the management objectives is evaluated as unknown and that area or complex is not used in the decision structure of the framework.

Within the geographic areas for which indicator variables are retained, all the available indicators are used to assess the indicator score. If an update value for an indicator variable is not available for the year of interest, the indicator variable is not used to quantify the indicator score for that area.

The West Greenland FWI was updated during the 2015 WGNAS meeting and an updated spreadsheet produced. Details are provided in Section 5.8 of the Working Group report (see above).

3.5.3 Framework of Indicators (FWI) for the Faroes Fishery

3.5.3.1 Background

The original FWI applied to the West Greenland fishery (ICES, 2007b) was not applicable for the Faroese fishery for a number of different reasons. Among these were the lack of quantitative catch advice, the absence of specific management objectives and a sharing agreement for this fishery and the fact that none of the available indicator datasets met the criteria for inclusion in the FWI.

In 2011, WGNAS re-evaluated the approach for developing a FWI for the Faroese fishery (ICES, 2011). Since over the available time-series the PFA estimates for the NEAC stock complexes have predominately remained above the SER, the Working Group suggested a different set of decision rules for this FWI. It was suggested that the status of stocks should be re-evaluated if the FWI suggests that the PFA estimates are deviating substantially from the median values from the forecast.

3.5.3.2 Description of the FWI

It was initially suggested that the 95% confidence interval range for the mean of the indicator prediction, relative to the median forecast value, be used to compute the decision thresholds for whether the indicator suggests a reassessment or not (ICES, 2011). The limits should be computed at the median values of the PFA forecasts in each of the years in multiyear advice. However, the 95% criterion was subsequently re-examined (ICES, 2012a) and it was recommended that the upper and lower 75% confidence limits of the individual predictions be used for comparison (Figure 3.5.3.2.1). WGNAS recognised that this was a relaxation of the decision rule suggested in 2011, and will lead to a larger interval, and thus a lower chance of a reassessment than the approach suggested in 2011. However, this was considered to be a more realistic confidence level given the relatively wide variability of the indicator datasets, and was also consistent with the approach adopted by NAC.

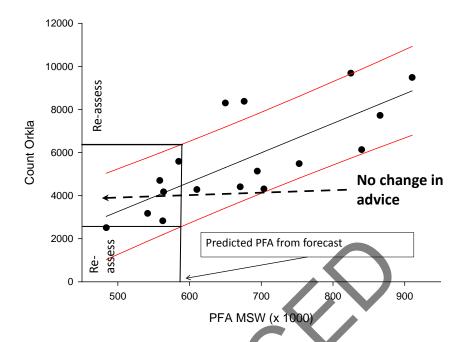


Figure 3.5.3.2.1. Example of how the reassessment intervals for the indicators are computed. The values of an indicator (counts) are plotted against the PFA. The regression line is shown in black and 75% confidence limits for the individual estimates are shown in red. From the forecasted PFA in the year in question the values of the indicator corresponding to the upper and lower 75% confidence interval are estimated. If the indicator value falls outside these limits a reassessment is recommended by this particular indicator.

When the stocks are divided into smaller management units, potential indicators for each management unit become relatively scarce. Therefore, the Working Group recommended that the FWI be regressed against the stock complexes that they belong to. For example, MSW indicators from Norway should be regressed against PFA MSW for northern NEAC.

In 2012, the FWI was applied as a two-tailed test (ICES, 2012a). However, it was subsequently agreed that, in the event of a closed fishery, the indicators should only be compared to the upper 75% confidence limit (i.e. a one-tailed test). This means that for a closed fishery, a reassessment is only triggered where the forecast appeared to be an underestimate and there may be a possibility of a harvest being denied. In the case of an open fishery they should be compared to both the upper and lower 75% confidence limits. In this case, if the FWIs suggest that the forecasted PFA is either an underestimation or an overestimation of the realised PFA in any of the four stock complexes, then this should trigger a reassessment.

WGNAS developed a FWI spreadsheet (ICES, 2011) to provide an automatic evaluation of the need for a reassessment once the new indicator values are available in January; this has been updated in subsequent years (ICES, 2012a; 2013). An example spreadsheet is provided at Figure 3.5.3.2.2.

The following summarizes the main steps performed by the spreadsheet following updating of the relevant data for the variable of interest by adding the latest year's number:

• Regression analysis with the dataset *x* to determine its power to predict PFA in the forecasted years.

- Calculation of the 75% confidence intervals of individual predictions of the regression for dataset *x*. An indicator value below the 75% individual confidence interval (CI) is interpreted as indicative of an overestimation of the PFA, while a point above the 75% individual confidence interval is interpreted as indicative of an underestimation of PFA.
- A dataset is considered informative and should be kept as an indicator in the FWI if the following conditions are met: sample size $(n) \ge 10$; $r^2 \ge 0.2$; dataset updated annually and new value available by January 15. Datasets that do not meet these criteria are discarded.
- Apply a binary score to each indicator value. Thus, for dataset *x*, if the current year's indicator value is outside the 75% individual regression point estimate CI (below or above) then that indicator receives a score of 1. If the indicator is within the 75% CI, then the indicator receives a score of 1. In the absence of an indicator datapoint for any year, a score of zero is applied. Whether the indicator value is above or below the upper and lower CI values is checked separately in two spreadsheet columns and a decision whether the indicator value is within the CI is assessed by combining the information in the two columns.
- Separate columns are used to sum the scores for all the indicator datasets within each stock complex. This is done separately for points that fall above the CI and those that fall below. In the case of a two-sided approach (open fishery), if the sum of these columns is ≥0, then the spreadsheet signals "REASSESS"; if the sum is <0, then it signals "No significant OVERestimation of PFA identified by indicators, do not reassess" for indicator values that fall below the CI, and "No significant UNDERestimation of PFA identified by indicators, do not reassess" for indicator values that are above the CI. In case of a one-sided approach (closed fishery), only underestimation will signal a "REASSESS".
- FWI results are generated for each stock complex (northern NEAC maturing and non-maturing, and southern NEAC maturing and non-maturing). A score of ≥0 for any of these stock complexes would signal a reassessment.

WGNAS reassessed the effects of applying stricter criteria than $r^2 \ge 0.2$ for inclusion of indicators in the FWI. As stricter criteria are used, the number of indicators included reduces rapidly. It was therefore concluded to keep the criterion of $r^2 \ge 0.2$ in order to obtain a sufficient number of indicators to be able to use the FWI even in the event of one or more indicators being unavailable by the time the FWI is applied each year. The r^2 value of 0.2 corresponds to a value slightly lower than what is considered to be a "large" effect size (r = 0.5, $r^2 = 0.25$) by Cohen (1988). Although a criterion of $r^2 \ge 0.2$ gives each indicator little predictive power alone (Prairie, 1996), the approach of using a suite of indicators is more similar to meta-analysis (Rosenthal, 1984) meaning that the outcome of the FWI is not dependent on the result of one indicator in isolation, but rather on the combined performance of the indicator set.

The Faroes FWI was updated during the 2015 WGNAS meeting and an updated spreadsheet produced. Details are provided in Section 3.7 of the Working Group report (see above). In evaluating all the time-series, it was noted that the lower 12.5 %

CL, which is used to determine which indicator values are outside the 75% CI on the lower side, was negative for some regression relationships for predicted PFA values in 2015 and 2016. Since this would invalidate the use of such indicators (they would not indicate that predicted PFA values are overestimates regardless of how small they are), an additional (fourth) criterion was established as a requirement for including time-series in the FWI. This requires that the lower 12.5% confidence limit for an indicator time-series should be positive for any values of PFA included in the FWI.

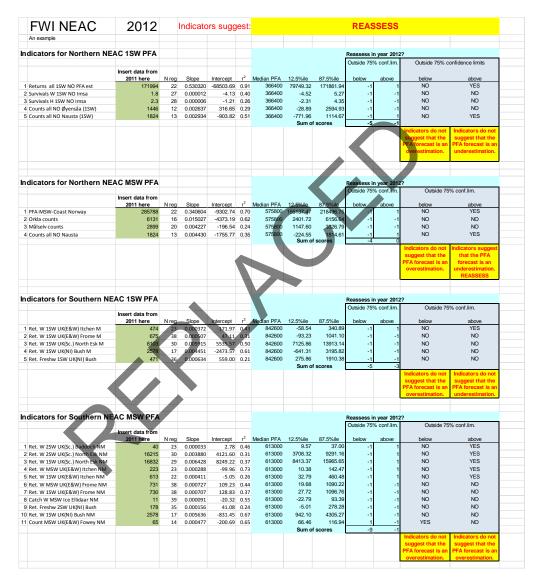


Figure 3.5.3.2.2. Output of the spreadsheet for the test of FWIs for NEAC for 2012 based on the values of the indicators from 2011. Because the indicators suggest that the forecast for northern NEAC MSW was an underestimate, the overall advice from the spreadsheet is reassess.

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Appendices to Stock Annex

Appendix 1 (a): Description of how catch-and-release mortality is incorporated in regional and national stock assessments

Commission			
Area	Country/Region	How it is used in regional and national assessments	Future developments / improvements
NAC	Canada-Quebec		New studies of the contribution of C&R fish to spawning success have been initiated. C&R monitoring is becoming more complete. Consideration will be given in the future to incorporating these losses in the returns and in the assessments based on angling catches, especially as reporting improves. No plans for further development.
	Newfoundland & Labrador	spawners. Spawning escapement is reduced by 5-15% (mean 10%) of the released catch.	
	Canada - Gulf	Assessments of spawners are adjusted by mortality rates of 3% to 6% of the total C&R estimates of small and large salmon. The rates vary by river according to angling seasons, and the occurence of other factors such as disease which can affect survival of salmon.	Catch and release mortality is known to be affected by the water temperatures when fish are angled. In some cases, angling fisheries are closed when water temperatures are high in the summer to reduce the losses of fish from C&R. Methods to determine catch and release numbers vary by over and in some cases, the number of released fish is estimated from returns and historical creel survey data. As the practice of C&R becomes more popular, estimation methods for C&R values will have to be revisited.
	Canada – Scotia/Fundy	Assessments are currently adjusted by 4% of the C&R fish to correct for C&R mortality.	Numbers of C&R fish are currently low (retention fisheries are closed). If C&R catches increase, further research on the correction factor would be warranted.
	USA	No correction for mortality due to C&R used in estimating spawner numbers. However, all fisheries have been closed and the number of fish caught relative to stock size is very small.	
NEAC	Russia	With increasing C&R the retained catch for similar effort is reduced. Therefore the exploitation rate for	
	Norway Sweden	retained fish is lower. The increase in C&R in recent years is incorporated into the national run- reconstruction model by reducing the exploitation rate	If C&R information is incorporated into formal assessments then muliple recaptures should be taken into account. C&R mortality should be incorporated into estimates of
	Iceland UK(Scotland)	value used in the model input. This is assessed qualitatively. No correction for increased C&R mortality is applied when estimating the spawning escapement.	spawning escapement.
•	Ireland	No correction for mortality due to C&R used in estimating spawner numbers or in the national run- reconstruction model.	Incorporation of formal method for estimating the effect of C&R on number of returning fish. Incorporation of C&R mortality in estimates of spawning escapement
	UK(England & Wales)	With increasing C&R the retained catch for similar effort is reduced. Therefore the exploitation rate for retained fish is lower. The increase in C&R in recent years is incorporated into the national runreconstruction model by reducing the exploitation rate value used in the model input. This is assessed qualitatively. 20% mortality of C&R fish used in assessing compliance with river-specific conservation	If C&R information is incorporated into formal assessments then multiple recaptures should be taken into account.
	UK(N. Ireland)	Returns are estimated by raising the reported net catch by exploitation rate. No correction for increased C&R mortality is applied when estimating the spawning escapement.	If C&R information is incorporated into formal assessments then multiple recaptures should be taken into account. C&R mortality should be incorporated into estimates of spawning escapement.
	Denmark	C&R rates recorded, but no national run-reconstruction assessment applied.	
	Finland	No record of C&R	If C&R information is collected, it should be incorporated into formal assessments and multiple recaptures should be

Appendix 1 (b): Description of how unreported catch is incorporated in regional, national and international stock assessments

Commission				
Area	Country/Region	How it is used in regional and national assessments	How used in international assessments	Future developments / improvements
NAC	Canada-Quebec	Unreported catches are based on historical estimates relative to stock size or are provided by field conservation and protection staff. Unreported catches when available are included in the regional assessment of returns and spawners.		
	Canada- Newfoundland & Labrador	Catch statistics include estimates of harvests by log book non-respondents. Therefore they are included in the regional assessments and the PFA estimate. No account is taken of illegal fisheries.		If unreported catch estimates were provided they could be incorporated in the regional assessments and in the
	Canada - Gulf	Unreported catches are sometimes provided by Conservation and Protection Personnel and are estimates of illegal fishing removals within specific regions. Unreported catches have not been used in the assessments of returns or spawners.	Unreported catches which occur in marine waters outside the jurisdiction of the regions are not included in the run reconstruction models.	continent estimates of abundance and spawners. Unreported catch may be accounted for in either the returns or the spawners, depending upon when and where the illegal activity occurs relative to the location and time of the
	Canada – Scotia/Fundy	No adjustment made, with the exception of the Saint John River where returns/spawners are adjusted for estimated bycatch and poaching. In other rivers where assessments directly quantify spawners, returns would be underestimated if catch is under reported.		assessment model.
	USA	Unreported catch is estimated to be zero and therefore has no effect on national assessments.		
NEAC	Russia			
	Finland			
	Norway			1
	Sweden		National estimates (which incorporate	Incorporate revised estimates of
	Iceland	Minimum and maximum estimates of the unreporting	unreported ctaches) are aggregated to	minimum and maximum estimates of
	Ireland	rate are used in deriving national PFA estimates from	provide PFA, return and spawner	unreporting rate as national estimates
	UK(Scotland)	the catch of 1SW & MSW salmon.	estimates for stock complexes.	are improved.
	UK(England & Wales)			
	UK(N. Ireland)			
	France			
	Denmark	No national assessment		
NEAC	Faroes	Not applicable	Assumed to be negligible unreported catch. Estimate of discard mortality for 1SW fish is incorporated in stock assessments.	Sampling programme if fishery resumes.
W. Greenland	W. Greenland	Not applicable	Unreported catch at West Greenland is incorporated in assessments for both the NAC and NEAC areas. Since 1993, this has been provided by the Greenlandic authorities. Prior to this time, no unreported catch component is included in the models.	Annual variation in unreported catch estimates would be incorporated into the model.

Appendix 2: Overview of current DCF and future data needs for Atlantic salmon assessment/ scientific advice

TYPE OF DATA	COLLECTED UNDER DCF	Available to WG	Reviewed and evaluated by WG	USED IN CURRENT ASSESSMENT MODELS	FUTURE PLANS	Notes
How to be filled	Yes/	Yes/	Yes/	Yes/	Keep as current DCF/ Improve sampling intensity/ No need to be	Free text
	No/	No/	No/	No/	collected/ (other free text)	
	Partially	Partially	Partially	Partially used		
Fleet capacity	No **	No *	No	No	No need to be collected	See 'Fishing gear and effort'
Fuel consumption	No **	No *	No	No	No need to be collected	Many salmon fisheries use unpowered vessels
Fishing gear and effort	Partially **	Partially	Partially	Partially, but information	Use for estimation of exploitation rates.	Data required for all relevant
				requested by NASCO	Improve coverage and sampling intensity in DC-MAP	areas/fisheries
Landings	Partially **	Yes	Yes	Yes	Improve coverage in DC-MAP	Data required on: catch in numbers and weights for recreational and commercial fisheries in rivers, estuaries and coastal waters.
Discards	No **	No*	No	No	No need to be collected	Not relevant to salmon except (historically) in Faroes fishery. NB: 'catch and release' fish are deliberately caught and so not classed as discards.
Recreational fisheries	Partially **	Yes	Yes	Yes	Improve coverage in DC-MAP	Extent of DCF coverage unclear. Complete catch data needed for all recreational fisheries (see 'Landings')
Catch & Release	No **	Partially	Partially	No - but data requested by NASCO	Include collection in DC-MAP	Data on numbers of fish caught and released required for all recreational fisheries

TYPE OF DATA	COLLECTED UNDER DCF	AVAILABLE TO WG	Reviewed and evaluated by WG	USED IN CURRENT ASSESSMENT MODELS	FUTURE PLANS	Notes
cpue dataseries	Partially **	Partially	Partially	Partially	Improve sampling intensity in DC-MAP	Data used to generate national inputs to models
Age composition	Partially ** Some ageing based on fish lengths or weights	Yes	Yes	Yes	Improve coverage and sampling intensity in DC-MAP	Extent of DCF coverage unclear; sampling intensities in other fisheries inappropriate to salmon
Wild/reared origin (scale reading)	No **	Partially from other sources	Partially	Partially used - information on farmed fish is requested by NASCO	Improve sampling intensity in DC-MAP	Extent of DCF coverage unclear
Length and weight-at- age	Partially **	Partially	Yes	Yes - but some ageing based on fish lengths or weights	Improve sampling coverage in DC-MAP	DCF does not cover all relevant areas/fisheries; sampling intensities inappropriate to salmon
Sex ratio	No **	Yes- from other sources	Partially	Yes	Modify sampling intensity in DC-MAP	Estimates required at national/regional level every five years
Maturity	Not known **	No*	No	No	No need to be collected – all returning adults are mature	DCF requires collection but extent of coverage unclear; data not required for assessments
Fecundity	No **	Yes	Partially	Yes	Include collection in DC-MAP	Estimates required at national/regional level every five years
Data processing industry	No **	No **	No	No	No need to be collected	Requirement not clear
Juvenile surveys (Electrofishing)	Partially ** but not requested for Atlantic salmon in DCF	Yes	Partially	Partially	Include collection in DC-MAP	Data used to develop reference points and confirm stock status. Also required for assessments under WFD

TYPE OF DATA	COLLECTED UNDER DCF	AVAILABLE TO WG	Reviewed and evaluated by WG	USED IN CURRENT ASSESSMENT MODELS	FUTURE PLANS	Notes
Adult census data (Counters, fish ladders, etc.)	Partially ** but not requested for Atlantic salmon in DCF	Yes	Partially	Yes	Include collection in DC-MAP	Counts required for ~one river in 30. Data required to provide exploitation rates for assessments
Index river data (Smolt & adult trapping; tagging programmes; etc.)	Partially ** but not requested for Atlantic salmon in DCF	Yes	Partially	Yes	Include collection in DC-MAP	Index rivers are identified by ICES. Data used to develop reference points and inputs to assessment models
Genetic data (for mixed- stock analysis)	No **	Partially	Partially - for some mixed- stock fisheries	Not currently	Include collection in DC-MAP - sampling in mixed-stock fisheries every five years	Genetic analysis is now advised to provide more reliable stock composition in mixed-stock fisheries
Economic data	Not known **	No *	No	No -but data are of use to NASCO		Collection of economic data would be useful to managers
Aquaculture data	Not known **	Partially - marine farm production collected	Yes	No - but information on farm production is requested by NASCO		Currently not required for freshwater

Add other data type to the cells with a light blue shading, if needed.

^{*} Not asked for by the ICES WGNAS.

^{**)} Not mandatory for some or all areas/stocks/fisheries under the current DCF.

Appendix 3: Input data for NEAC Pre Fishery Abundance analysis using Monte Carlo simulation

Finland

arlo sir	nulation]									
Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty i
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitatio
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSV
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmo
1971	8422	8538	35	5	35	5	50	10	55	1
1972	32789	8950	35	5	35	5	50	10	55	1
1973	15261	14402	35	5	35	5	50	10	55	1
1974	21057	24508	35	5	35	5	50	10	55	1
1975	25242	31347	35	5	35	5	50	10	55	1
1976	23000	24561	35	5	35	5	50	10	55	1
1977	12958	17035	35	5	35	5	50	10	55	1
1978	12338	8670	35	5	35	5	50	10	55	1
1979	11071	7078	35	5	35	5	50	10	45	1
1980	10097	7994	25	5	25	5	50	10	45	1
1981	9049	9476	25	5	25	5	50	10	45	1
1982	5379	12628	25	5	25	5	50	10	45	1
1983	13156	14013	25	5	25	5	50	10	45	1
1984	14371	11718	25	5	25	5	50	10	45	1
1985	19058	11299	25	5	25	.5	50	10	45	1
1986	15005	9320	25	5	25	5	50	10	45	1
1987	18151	12208	25	5	25	5	50	10	45	1
1988	10676	8631	25	5	25	- 5	50	10	45	1
1989	27956	10337	25	5	25	5	60	10	55	1
1990	27955	11423	25	5	25	5	60	10	55	1
1991	27513	15287	25	5	25	5		10	55	1
1992	38843	14826	25	5	25	5	60	10	55	1
1993	26195	15517	25	5	25	5	60	10	55	1
1994	14555	14621	25	5	25	5	60	10	55	1
1995	14525	9625	25	5	25	5	60	10	55	1
1996	20466	8079	25	5	25	5	50	10	45	1
1997	18621	9764	25	5	25	5	50	10	45	1
1998	23336	9307	25	5	25	5	50	10	45	1
1999	37495	11071	25	5	25	5	60	10	50	1
2000	40730	21088	25	5	25	5	60	10	50	1
2001	29501	28112	25	5	25	5	60	10	50	1
2002	16721	24642	25	5	25	5	50	10	50	1
2003	16497	17751	25	5	25	5	50	10	50	1
2004	7002	8062	25	5	25	5	50	10	50	1
2005	15366	6685	25	5	25	5	50	10	50	1
2006	26916	10533	25	5	25	5	50	10	50	1
2007	7862	15269	25	5	25	5	50	10	50	1
2008	8481	15355	25	5	25	5	50	10	50	1
2009	15042	6587	25	5	25	5	50	10	50	1
2010	12085	10590	25	5	25	5	50	10	50	1
2010	13727	8152	25	5	25	5	50	10	50	1
2011	23764	9851	25	5	25	5	50	10	50	1
2012	13724	9494	25	5	25	5	50	10	50	1
2013	19495	10302	25	5	25	5	50	10	50	1
2014	12127	9905	25	5	25	5	50	10	50	1

France

	ulation]									
Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
· cu.	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
	50	501111011	salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmor
1971	1740	4060		NA	NA	NA	3.5	1.5	37.5	12.5
1972	3480	8120		NA	NA	NA	3.5	1.5	37.5	12.5
1973	2130	4970		NA	NA	NA	3.5	1.5	37.5	12.5
1974	990	2310		NA	NA	NA	3.5	1.5	37.5	12.5
1975	1980	4620		NA	NA	NA	3.5	1.5	37.5	12.5
1976	1820	3380		NA	NA	NA	3.5	1.5	37.5	12.5
1977	1400	2600		NA	NA	NA	3.5	1.5	37.5	12.5
1978	1435	2665		NA	NA	NA	3.5	1.5	37.5	12.5
1979	1645	3055		NA	NA	NA	3.5	1.5	37.5	12.5
1980	3430	6370		NA	NA	NA	3.5	1.5	37.5	12.5
1981	2720	4080		NA	NA	NA	3.5	1.5	37.3	12.5
1982	1680	2520		NA	NA	NA	3.5	1.5	35	15
1983	1800	2700		NA	NA	NA	3.5	1.5	35	15
1984	2960	4440		NA	NA	NA	3.5	1.5	35	15
1985	1100	3330		NA	NA	NA	3.5	1.5	35	15
1986	3400	3400		NA	NA	NA	7	5	35	15
1987	6013	1806		NA	NA	NA	7	5	35	15
1988	2063	4964		NA	NA	NA	7	5	35	15
1989	1124	2282		NA	NA	NA	7	5	35	15
1990	1886	2332		NA	NA	NA	7	5	35	15
1991	1362	2125		NA	NA	NA	7	5	35	15
1992	2490	2671		NA	NA	NA	7	5	35	15
1993	3581	1254		NA	NA	NA .	7	5	35	15
1994	2810	2290		NA	NA	NA NA	7	5	30	10
1995	1669	1095		NA	NA	NA NA	12.5	7.5	30	10
1996	2063	1943		NA	NA	NA NA	12.5	7.5	30	10
1997	1060	1001		NA	NA	NA	12.5	7.5	30	10
1998	2065	846		NA	NA	NA	12.5	7.5	30	10
1999	690	1831		NA	NA	NA	12.5	7.5	30	10
2000	1792	1277		NA	NA	NA	12.5	7.5	30	10
2001	1544	1489		NA	NA 👝	NA	12.5	7.5	30	10
2001	2423	1065	30	10		7.5	12.5	7.5	30	10
2002	1598	1540	30	10		7.5	12.5	7.5	30	10
2003	1927	2880	30	10		7.5	12.5	7.5	30	10
2004	1256	1771	30	10		7.5	12.5	7.5	30	10
2005	1763	1771	30	10		7.5	12.5	7.5	30	10
2007	1378	1685	30	10	22.5	7.5	12.5	7.5	30	10
2007	1365	1865	30			7.5	12.5	7.5	30	10
2009	389	863	30	10		7.5	12.5	7.5	30	10
2009	1313	711	30	10		7.5	12.5	7.5	30	10
2010	899	1998	30	10	_	7.5	12.5	7.5	30	10
2011	974	1585	30	10		7.5	12.5	7.5	30	10
2012	1371	1632	30	10		7.5	12.5	7.5	30	1
2013	1217	2027	30	10		7.5	12.5	7.5	30	10
2014	1124	2027	30	10		7.5	12.5	7.5	30	10

Iceland (South and West)

sed in	Monte Carlo s	imulation]								
Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmor
1971	30618	16749	2	1	2	1	50	10	70	10
1972	24832	25733	2	1	2	1	50	10	70	10
1973	26624	23183	2	1	2	1	50	10	70	10
1974	18975	20017	2	1	2	1	50	10	70	10
1975	29428	21266	2	1	2	1	50	10	70	10
1976	23233	18379	2	1	2	1	50	10	70	10
1977	23802	17919	2	1	2	1	50	10	70	10
1978	31199	23182	2	1	2	1	50	10	70	10
1979	28790	14840	2	1	2	1	50	10	70	10
1980	13073	20855	2	1	2	1	50	10	70	10
1981	16890	13919	2	1	2	1	50	10	70	10
1982	17331	9826	2	1	2	1	50	10	70	10
1983	21923	16423	2	1	2	1	50	10	70	10
1984	13476	13923	2	1	2	1	50	10	70	10
1985	21822	10097	2	1	2	1	50	10	70	10
1986	35891	8423	2	1	2	1	50	10	70	10
1987	22302	7480	2	1	2	1	50	10	70	10
1988	40028	8523	2	1	2	1	50	10	70	10
1989	22377	7607	2	1	2	1	50	10	70	10
1990	20584	7548	2	1	2	1	50	10	70	10
1991	22711	7519	2	1	2	1	50	10	70	10
1992	26006	8479	2	1	2	1	50	10	70	10
1993	25479	4155	2	1	2		50	10	70	10
1994	20985	6736	2	1	2	1	50	10	70	10
1995	25371	6777	4	1	4	1	50	10	70	10
1996	21913	4364	4	1	4	1	50	10	70	10
1997	16007	4910	4	1	4	1	50	10	70	10
1998	21900	3037	4	1	4	1	50	10	70	10
1999	17448	5757	4	1	4	1	49	10	68	10
2000	15502	1519	4	1	4	1	49	10	66	10
2001	13586	2707	4	1	4	1	48	10	67	10
2002	16952	2845	4	1	4	1	48	10	65	10
2003	20271	4751	4	1	4	1	48	10	68	10
2004	20319	3784	4	1	4	1	48	10	67	10
2005	29969	3241	4	1	4	1	48	10	65	10
2006	21153	2689	4	1	4	1	48	10	65	10
2007	23728	1679	4	1	4	1	47	9	66	10
2008	28774	1659	4	1	4	1	47	10	57	10
2009	33190	2838	4	1	4	1	48	10	63	10
2010	33318	6061	4	1	4	1	47	10	65	10
2011	23436	2934	. 4	1	4	1	47	10	62	10
2012	13312	1429	4	1	4	1	47	10	53	10
2013	39637	4105	4	1	4	1	47	10	55	10
2014	9551	2281	4	1	4	1	46	10	50	10
2015	26079	1816	4	. 1	4	1	46	10	51	10

Iceland (North and East)

		imulation]								
Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty i
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitatio
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmo
1971	4610	6625	2	1	2	1	50	10	70	1
1972	4223	10337	2	1	2	1	50	10	70	1
1973	5060	9672	2	1	2	1	50	10	70	1
1974	5047	9176	2	1	2	1	50	10	70	1
1975	6152	10136	2	1	2	1	50	10	70	1
1976	6184	8350	2	1	2	1	50	10	70	1
1977	8597	11631	2	1	2	1	50	10	70	1
1978	8739	14998	2	1	2	1	50	10	70	1
1979	8363	9897	2	1	2	1	50	10	70	1
1980	1268	13784	2	1	2	1	50	10	70	1
1981	6528	4827	2	1	2	1	50	10	70	1
1982	3007	5539	2	1	2	1	50	10	70	1
1983	4437	4224	2	1	2	1	50	10	70	1
1984	1611	5447	2	1	2	1	50	10	70	1
1985	11116	3511	2	1	2	1	50	10	70	1
1986	13827	9569	2	1	2	1	50	10	70	1
1987	8145	9908	2	1	2	1	50	10	70	1
1988	11775	6381	2	1	2	1	50	10	70	1
1989	6342	5414	2	1	2	1	50	10	70	1
1990	4752	5709	2	1	2	1	50	10	70	1
1991	6900	3965	2	1	2	1	50	10	70	1
1992	12996	5903	2	1	2	1	50	10	70	1
1993	10689	6672	2	1	2	1	50	10	70	1
1994	3414	5656	2	1	2		50	10	70	1
1995	8776	3511	4	1	4	1	50	10	70	1
1996	4681	4605	4	1	4	1	50	10	70	1
1997	6406	2594	4	1	4	1	50	10	70	1
1998	10905	3780	4	1	4	1	50	10	70	1
1999	5326	4030	4	1	4	1	48	10	65	1
2000	5595	2324	4	1	4	1	48	10	64	1
2001	4976	2587	4	1	4	1	47	10	62	1
2002	8437	2366	4	1	4	1	46	10	60	1
2003	4478	2194	4	1	4	1	46	10	53	1
2004	11823	2239	4	1	4	1	45	10	55	1
2005	10297	2726	4	1	4	1	44	10	54	1
2006	11082	2179	4	1	4	1	45	10	45	1
2007	8046	1672	4	1	4	1	44	10	36	1
2008	7021	2693	4	1	4	1	42	10	45	1
2009	10779	1735	4	1	4	1	40	10	36	1
2010	8621	2602	4	1	4	1	40	10	38	1
2011	6759	2596	4	1	4	1	38	10	34	1
2012	3699	1419	4	1	4	1	40	10	33	1
2013	8375	1528	. 4	1	4	1	38	10	31	1
2014	3953	1778	4	1	4	1	38	10	30	1
2015	10070	1775	4	. 1	4	1	34	10	30	1

Ireland

_																		
Year	Declared	Declared		Uncertainty in		Uncertainty in		Uncertainty in		Uncertainty in	Declared net				1SW salmon in	MSW		
	catch 1SW	catch MSW		% unreported		% unreported		exploitation	exploitation	exploitation	catch 1SW				Small rivers			
	salmon	salmon		catch of 1SW			` '	٠, ,	rate (%) - MSW salmon	rate (%) -	salmon	salmon	salmon	MSW salmon		Small rivers		
			salmon	salmon	salmon	salmon	salmon	saimon	IVISW saimon	IVISW Salmon				saimon			rivers	river
1971	409965	46594	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5							NA	NA
1972	437089	49863	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5							NA	NA
1973	476131	54008	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5							NA	NA
1974	542124	60976	37.5	7.5	37.5	7.5	62.5	12.5	47.5 47.5	12.5		_					NA	NA
1975 1976	598524 407018	68260 47358	37.5 37.5	7.5 7.5	37.5 37.5	7.5 7.5	62.5 62.5	12.5 12.5	47.5 47.5	12.5 l	$\overline{}$	_					NA NA	NA NA
1977	351745	41256	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5							NA	NA
1978	307569	35708	37.5	7.5	37.5	7.5	62.5		47.5	12.5	_						NA	NA
1979	282700	32144	37.5	7.5	37.5	7.5	62.5		47.5	12.5							NA	NA
1980	215116	35447	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5	_						NA	NA
1981	137366	26101	37.5	7.5	37.5	7.5	75.74	11.36	47.5								NA	NA
1982	269847	11754	37.5	7.5	37.5	7.5	71.86	10.78	36.665	8.325	_						NA	NA
1983	437751	26479	37.5	7.5	37.5	7.5	66.05	9.91	40.1	7.5	NA	NA	NA	NA	NA	NA	NA	NA
1984	224872	20685	37.5	7.5	37.5	7.5	64.595	9.685	43,51	6.49	AV	NA	NA	NA	NA	NA	NA	NA
1985	430315	18830	37.5	7.5	37.5	7.5	74.575	11,185	36.1	3.35	NA	NA	NA	NA	NA	NA	NA	NA
1986	443701	27111	37.5	7.5	37.5	7.5	68.705	10.305	45.975	9.025	AV	NA	NA	NA	NA	NA	NA	NA
1987	324709	26301	30	10	30	10	69.81	10.47	32.18	4.68	AV	NA	NA	NA	NA	NA	NA	NA
1988	391475	22067	30	10	30	10	62.035	9.305	37.425	5.575							NA	NA
1989	297797	25447	30	10	30	10	65.705	9.855	47.175	8.825 I							NA	NA
1990	172098	15549	30	10	30	10	60.73	9.11	59.925	6.075							NA	NA
1991	120408	10334	30	10	30	10	59.47	8.92		3.5							NA	NA
1992	182255	15456	30	10	30	10	62.055	9.305	51.46	3.8							NA	NA
1993	150274	13156	25	10	25	10	58.645	8.795	42	18							NA	NA
1994	234126	20506	25	10	25	10	71.41	10.71	40.53	2.47							NA	NA
1995 1996	232480 203920	20454 18021	25 25	10 10	25 25	10 10	63.46 59.885	9.52 8.985	41.825 55.105	1.175 3.175							NA NA	NA NA
1996	170774	14724	25	10	15	10	50.105	7.515	30.755	12.245							NA	NA
1998	191868	17269	25	10	15		53.72	8.06	61.86	1.39							NA	NA
1999	158818	14801	25	10	15	5			34.145	18.145							NA	NA
2000	199827	16848	25	10	15		43.235	6.485	30.995	4.485							NA	NA
2001	218715	18436	7.5	2.5	7.5		48	7.2	35	8 1							NA	NA
2002	198719	16702	7.5	2.5	7.5	2.5	49.89	7.48	27.5	7.5							NA	NA
2003	161270	13745	7.5	2.5	7.5	2.5	41.325	6.195	21.5	5.5 [NA	NA	NA	NA	NA	NA	NA	NA
2004	142251	12299	7.5	2.5	7.5	2.5	49.5	7.5	35	8 1	AV	NA	NA	NA	NA	NA	NA	NA
2005	127371	10716	7.5	2.5	7.5	2.5	44.5	6.5	23.5	3.5	AV	NA	NA	NA	NA	NA	NA	NA
2006	101938	9740	7.5	2.5	7.5	2.5	46.5	6.5	29.5	13.5 I	AV	NA	NA	NA	NA	NA	NA	NA
2007	17863	2867	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	8177	666		988	0	-		
2008	31843	3935	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	8233	670		1492	0	-		
2009	24268	4675	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	6248	509			0			
2010	32981	4497	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	13093	1066			0	-		
2011	28105	4889	7.5	2.5	7.5		15.5	8.4	23.9	9.1	11071				0			
2012	29979	4197	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	9542	777			0			
2013	24029	4831	7.5	2.5	7.5		15.5	8.4	23.9	9.1	13378				0			
2014	15328	2522	7.5 7.5	2.5 2.5	7.5 7.5	2.5 2.5	15.5	8.4 8.4	23.9 23.9	9.1 9.1	8903 7396	667 295		1430	0	-		

Norway (Southeast)

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (SOUTH-EAST). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation] Declared Declared Estimated % Uncertainty in Estimated % Uncertainty in Estimated Uncertainty in Estimated Uncertainty in unreported % unreported winreported catch of 1SW catch of MSW catch of catch 1SW catch MSW salmon salmon salmon salmon salmon 1971 NA 1972 NA 1973 NA 1974 NA NA NΑ NΑ NA NΑ NΑ NA NA NA 1975 NA 1976 NA 1977 NA NA NΑ NΑ NA NΑ NΑ NA NA NA 1978 NA 1979 NA NA NA NΑ NA NA NA NA NA 1980 NA 1981 NA NA NA NA NA NA NA NA NA 1982 NA 50 50 70 65 12254 50 50 60 50 50 60 65 1995 11630 10 40 10 60 10 60 2006 30 10 60 60 10

Norway (Southwest)

Year	Declared catch 1SW salmon	Declared catch MSW salmon	Estimated % unreported catch of 1SW	Uncertainty in % unreported catch of 1SW	Estimated % unreported catch of MSW	Uncertainty in % unreported catch of MSW	Estimated exploitation rate (%) - 1SW	Uncertainty in exploitation rate (%) - 1SW	Estimated exploitation rate (%) - MSW	Uncertainty in exploitation rate (%) - MSW
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971		NA	NA		NA	NA	NA	NA	NA	NA
1972		NA	NA		NA	NA	NA	NA	NA	NA
1973		NA	NA		NA	NA	NA	NA	NA	NA
1974		NA	NA		NA	NA	NA	NA	NA	NA
1975		NA	NA		NA	NA	NA	NA	NA	NA
1976		NA	NA		NA	NA	NA	NA	NA	NA
1977		NA	NA		NA	NA	NA	NA	NA	NA
1978		NA	NA		NA	NA	NA	NA	NA	NA
1979			NA		NA	NA	NA	NA	NA	NA
1980	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
1981	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
1982	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
1983	31845	28601	50	10	50	10	80	10	80	10
1984	23428	27641	50	10	50	10	80	10	80	10
1985	29857	25515	50	10	50	10	80	10	80	10
1986	29894	30769	50	10	50	10	80	10	80	10
1987	30005	26623	50	10	50	10	80	10	80	10
1988	36976	28255	50	10	50	10	80	10	80	10
1989	19183	13041	50	10	50	10	70	10	65	10
1990	18490	14423	50	10	50	10	70	10	65	10
1991	9759	8323	50	10	50	10		10	65	10
1992	6448	8832	50	10	50	10		10	65	10
1993	11433	10239	40	10	40	10		10	65	10
1994	18597	10961	40	10	40	10		10	65	10
1995	10863	13122	40	10	40	10		10	65	10
1996	7048	12546	40	10	40	10	70	10	65	10
1997	10279	7194	35	10	35	10		10	60	10
1998	5726	6583	35	10	35	10		10	60	10
1999	7357	3219	35	10	35	10		10	60	10
2000	11538	7961	35	10	35	10		10	60	10
2001	12109	10716	35	10	35	10	_	10	60	10
2001	6000	7145	35	10	35	10		10	60	10
2002	8269	7602	30	10	30			10	60	10
2003	7180	6420	30	10	30			10	60	10
2005	10370	7334	30	10	30	10	60	10	60	10
2005	5173	9381	30	10	30	10		10	60	10
2007	2630	6011	30	10	30	10		10	60	10
2007	3143	4807	30	10	30	10		10	50	10
2008	3069	3792	30		30	10		10	50	10
	3450	2447	30	10		10		10	35	
2010 2011						10		10	35	10
	2888	4409	30			10		10		10
2012	4171	5733	30	10	30				30	10
2013	3111	3581	30	10	30	10		10	30	10
2014 2015	3029 4721	2717 3953	30	10 10	30 30	10 10		10 10	25 30	10

Mid-Norway

_										
Year		Declared	Estimated %	Uncertainty in		Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW		catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971			NA	NA	NA	NA	NA	NA		NA
1972				NA	NA	NA	NA	NA		NA
1973			NA	NA	NA	NA	NA	NA		NA
1974			NA	NA	NA	NA	NA	NA		NA
1975			NA	NA	NA	NA	NA	NA		NA
1976			NA	NA	NA	NA	NA	NA		NA
1977			NA	NA	NA	NA	NA	NA		NA
1978			NA	NA	NA	NA	NA	NA		NA
1979			NA	NA	NA	NA	NA	NA		NA
1980			NA	NA	NA	NA	NA	NA		NA
1981			NA	NA	NA	NA	NA	NA		NA
1982	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
1983	121221	74648	50	10		10	75		75	10
1984	94373	67639		10		10	75	10	75	10
1985	114613	56641	50	10	50	10	75	10	75	10
1986	106921	77225	50	10	50	10	75	10	75	10
1987	83669	62216	50	10	50	10	75	10	75	10
1988	80111	45609	50	10	50	10	75	10	75	10
1989	94897	30862	50	10	50	10		10	65	10
1990	78888	40174	50	10	50	10	65	10	65	10
1991	67370	30087	50	10	50	10	65	10	65	10
1992	51463	33092	50	10	50	10	65	10	65	10
1993	58326	28184	40	10	40	10	65	10	65	10
1994	113427	33520	40	10	40	10	65	10	65	10
1995	57813	42696	40	10	40	10	65	10	65	10
1996	28925	31613	40	10	40	10	65	10	65	10
1997	43127	20565	35	10	35	10	60	10	60	10
1998	63497	26817	35	10	35	10	60	10	60	10
1999	60689	28792	35	10	35	10	60	10	60	10
2000	109278	42452	35	10	35	10	60	10	60	10
2001	88096	52031	35	10	35	10	60	10	60	10
2002	42669	52774	35	10	35	10	60	10	60	10
2003	91118	46963	30	10	30	10	60	10	60	10
2004		49760	30	10	30	10	60	10	60	10
2005	63749	37941	30	10	30	10	60	10	60	10
2006		47691	30	10		10	60	10	60	10
2007		33106		10			60	10	60	10
2008	31936	34869	30	10	30	10	55	10	45	10
2009	26267	30715		10		10			45	10
2010		30524		10		10		10	45	10
2011	20932	37272	_	10	_	10		10	45	10
2012		28265		10		10		10	45	10
2013		17727	30			10	45	10	40	10
2014		14199	30	10		10	35		32	10
2015		30457	30	10		10		10	40	10

Norway North

Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971				NA	NA	NA	NA		NA	NA
1972				NA	NA	NA	NA	NA	NA	NA
1973	NA			NA	NA	NA	NA	NA	NA	NA
1974	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
1975			NA	NA	NA	NA	NA	NA	NA	NA
1976			NA	NA	NA	NA	NA	NA	NA	NA
1977	NA		NA	NA	NA	NA	NA	NA	NA	NA
1978			NA	NA	NA	NA	NA	NA	NA	NA
1979				NA	NA	NA	NA	NA	NA	NA
1980	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
1981			NA	NA	NA	NA	NA	NA	NA	NA
1982				NA	NA	NA	NA	NA	NA	NA
1983		49412.81776		10	50	10	80	10	80	10
1984				10	50		80	10	80	10
1985	118840.8653		50	10	50		80	10	80	10
1986			50	10	50	10	80	10	80	10
1987	72369.54095			10	50		80	10	80	10
1988		32139.51427		10	50		80	10	80	10
1989		13934.22823		10	50		70	10	70	10
1990		17320.80962		10	50	10	70	10	70	10
1991		21788.71036		10	50	10	70	10	70	10
1992	34592.56357	19265.3515	50	10	50	10	70	10	70	10
1993	51439.62347	39014.46832	40	10	40	10	70	10	70	10
1994	37489.00853	33411.4443	40	10	40		70	10	70	10
1995	36283.47628	26036.60329	40	10	40	10	70	10	70	10
1996	40792.02271	36636.04195	40	10	40	10	70	10	70	10
1997	39929.57821	30114.87466	35	10	35	10	70	10	70	10
1998	46644.59827	34805.94209	35	10	35	10	70	10	70	10
1999	46393.95449	46744.40937	35	10	35	10	70	10	70	10
2000	61854.4489	51568.94298	35	10	35	10	70	10	70	10
2001	46331	54023	35	10	35	10	70	10	70	10
2002	38101	43100	35	10	35	10	70	10	70	10
2003	44947	35972	30	10	30	10	70	10	70	10
2004	34640	28077	30	10	30	10	70	10	70	10
2005	45530	33334	30	10	30	10	70	10	70	10
2006	48688	39508	30	10	30	10	70	10	70	10
2007	28748	44550	30	10	30	10	70	10	70	10
2008	34338	40553	30	10	30	10	65	10	65	10
2009	22511	28241	30	10	30	10	65	10	65	10
2010	29836	28611	30	10	30	10	65	10	55	10
2011	26813	27233	_	10	30	10	65	10	55	10
2012	28289	28000	30	10	30	10	65	10	55	10
2013	20021	24689	30	10	30	10	65	10	55	10
2014	35171	23816	30	10	30	10	65	10	55	10
	25426	23890	30	10	30	10	65	10	55	10

Russia (Archangelsk and Karelia)

Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	134	16592	10	5	10	5	60	20	60	20
1972	116	14434	10	5	10	5	60	20	60	20
1973	169	20924	10	5	10	5	60	20	60	20
1974	170	21137	10	5	10	5	60	20	60	20
1975	140	17398	10	5	10	5	60	20	60	20
1976	111	13781	10	5	10	5	60	20	60	20
1977	78	9722	10	5	10	5	60	20	60	20
1978	82	10134	10	5	10	5	60	20	60	20
1979	112	13903	10	5	10	5	60	20	60	20
1980	156	19397	10	5	10	5	60	20	60	20
1981	68	8394	10	5	10	5	60	20	60	20
1982	71	8797	10	5	10	5	60	20	60	20
1983	48	11938	10	5	10	5	60	20	60	20
1984	21	10680	10	5	10	5	60	20	60	20
1985	454	11183	10	5	10	5	60	20	60	20
1986	12	12291	10	5	10	5	60	20	60	20
1987	647	8734	10	5	10	5	60	20	60	20
1988	224	9978	10	5	10	5	60	20	60	20
1989	989	10245	10	5	10	5	60	20	60	20
1990	1418	8429	15	5	15	5	60	20	60	20
1991	421	8725	20	5	20	5	60	20	60	20
1992	1031	3949	25	5	25	5	60	20	60	20
1993	196	4251	30	5	30	5	60	20	60	20
1994	334	5631	35	5	35	5	60	20	60	20
1995	386	5214	45	5	45	5	60	20	60	20
1996	231	3753	55	5	55	5	60	20	60	20
1997	721	3351	55	5	55	5	60	20	60	20
1998	585	4208	55	5	55	5	60	20	60	20
1999	299	3101	55	5	55	5	60	20	60	20
2000	514	3382	55	5	55	5	60	20	60	20
2001	363	2348	55	5	55	5	60	20	60	20
2002	1676	2439	55	5	55	-5	60	20	60	20
2003	893	2041	55	5	55 55	5 5	60	20	60	20
2004	990	3761	55	5			60	20	60	20
2005	1349	4915	55	5	55	5	60	20	60	20
2006	2183	2841	55	5	55	5	60	20	60	20
2007	1618	2621	55	5	55	5	60	20	60	20
2008	332	2496	55	5	55	5	60	20	60	20
2009	252	2214	55	5	55	5	60	20	60	20
2010	397	3823	55	5	55	5	60	20	60	20
2011	313	2585	55	5	55	5	60	20	60	20
2012	1332	2446	55	5	55	5	60	20	60	20
2013	2296 2084	3480 3463	55 55	5 5	55 55	5 5	60 60	20 20	60 60	20
2014										

Russia (Kola Peninsula: Barents Sea Basin)

		imulation]								
Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSV
4074	4000	5070	salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmor
1971	4892	5979	15	5	15	5	45	5	45	
1972	7978	9750	15	5 5	15	5 5	45	5 5	45	
1973	9376	11460	15		15		40		40	
1974	12794	15638	15	5	15	5	40	5	40	
1975	13872	13872	15	5	15	5	45	5	45	
1976	11493	14048	15	5	15	5	55	5	55	
1977	7257	8253	15	5	15	5	50	5	50	
1978	7106	7113	15	5	15	5	55	5	55	
1979	6707	3141	15	5	15	5	40	5	40	
1980	6621	5216	15	5	15	5	40	5	40	
1981	4547	5973	15	5	15	5	40	5	40	
1982	5159	4798	15	5	15	5	35	5	35	
1983	8504	9943	15	5	15	5	35	5	35	!
1984	9453	12601	15	5	15	5	35	5	35	
1985	6774	7877	15	5	15	5	35	5	35	
1986	10147	5352	15	5	15	5	40	5	40	
1987	8560	5149	15	5	15	5	40	5	40	
1988	6644	3655	15	5	15	5	35	5	35	
1989	13424	6787	15	5	15	5	40	5	40	
1990	16038	8234	15	5	15	5	40	5	40	
1991	4550	7568	15	5	15	5	30	5	30	
1992	11394	7109	15	5	15	5	30	5	30	
1993	8642	5690	15	5	15	5	30	5	30	!
1994	6101	4632	15	5	15	5	30	5	30	
1995	6318	3693	15	5	15	5	30	5	30	
1996	6815	1701	20	5	20	5	25	5	25	į
1997	3564	867	25	5	25	5	15	5	15	
1998	1854	280	35	5	35	5	12.5	2.5	12.5	2.5
1999	1510	424	40	5	40	5	7.5	2.5	7.5	2.5
2000	805	323	50	5	50	5	6	2	6	
2001	591	241	60	5	60	5	3.5	1.5	3.5	1.5
2002	1436	2478	50	10	50	_10	10	5	20	
2003	1938	1095	50	10	50	10	10	5	20	!
2004	1095	850	50	10	50	10	10	5	20	!
2005	859	426	60	10	60	10	10	5	20	!
2006	1372	844	60	10	60	10	10	5	20	!
2007	784	707	60	10	60	10	10	5	20	
2008	1446	997	60	10	60	10	15	5	20	!
2009	2882	1080	60	10	60	10	15	5	20	!
2010	3884	1486	60	10	60	10	20	5	25	
2011	3861	1407	60	10	60	10	20	5	25	
2012	2708	1027	60	10	60	10	20	5	25	
2013	939	904	60	10	60	10	20	5	25	
2014	969	789	60	10	60	10	20	5	25	
2015	727	494	60	10	60	10	20	5	25	

Russia (Kola Peninsula: White Sea Basin)

Year	Declared	Declared	Catch 1SW	Catch MSW	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	following-year	following-year	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	spawners	spawners	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
					salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	67845	29077	0	0	3	2	3	2	50	10	60	10
1972	45837	19644	0	0	3	2	3	2	50	10	60	10
1973	68684	29436	0	0	3	2	3	2	50	10	60	10
1974	63892	27382	0	0	3	2	3	2	50	10	60	10
1975	109038	46730	0	0	3	2	3	2	50	10	60	10
1976	76281	41075	0	0	3	2	3	2	50	10		10
1977	47943	32392	0	0	3	2	3	2	50	10	60	10
1978	49291	17307	0	0	3	2	3	2	50	10	60	10
1979	69511	21369	0	0	3	2	3	2	50	10		10
1980	46037	23241	0	0	3	2	3	2	50	10		10
1981	40172	12747	0	0	3	2	3	2	50	10		10
1982	32619	14840	0	0	3	2	3	2	50	10	60	10
1983	54217	20840	0	0	3	2	3	2	50	10		10
1984	56786	16893	0	0	3	2	3	2	50	10	60	10
1985	87274	16876	0	0	3	2	3	2	50	10		10
1986	72102	17681	0	0	3	2	3	2	50	10	60	10
1987	79639	12501	0	0	3	2	3	2	50	10		10
1988	44813	18777	0	0	3	2	3	2	45	5		5
1989	53293	11448	0	0	7.5	2.5	7.5	2.5	45	5		5
1990	44409	11152	0	0	12.5	2.5	12.5	2.5	45	5		5
1991	31978	6263	0	0	17.5	2.5	17.5	2.5	35	5		5
1992	23827	3680	0	0	22.5	2.5	22.5	2.5	25	5		5
1993	20987	5552	0	0	25	5	25	5	25	5		5
1994	25178	3680	0	0	30	5	30	5	25			
1995	19381	2847	0	0	35	5	35	5	25	5		
1996	27097	2710	0	0	35	5	35	5	25	5		
1997	27695	2085	0	0	35	5	35	5	25	5		
1998	32693	1963	0	0	35	5	35	5	25	5		5
1999	22330	2841	0	0	35	5	35	5	25	5		
2000	26376	4396	0	0	35	5	35	5	25	5		
2001	20483	3959	0	0	35	5	35	5	15	5		5
2002	19174	3937	0	0	35	5	35	5	15	5		
2003	15687	3734	0	0	35	5	25	5	15	5		
2004	10947	1990	0	0	35	5	35	5	15	5		
2005	13172	2388	1212	878	35	5	35	5	15	5		
2006	15004	2071	3852	399	35	5	35	5	15	5		
2007	7807	1404	2264	852	35	5	35	5	15	5		
2008	8447	4711	3175	832	35	5	35	5	15	5		
2009	5351	3105	5130	1710	35	5	35	5	15	5		
2010	6731	4158	3684	1228	35	5	35	5	15	5		
2011	7363	4325	3082	1027	35		35	5	15	5		
2012	10398	1431	2267	756	35	5	35	5	15	5		
2013	8986	1660	2203	734	35	5	35	5	15	5		
2014	8593	1674	3307	1102	35	5	35	5	15	5		5
2015	9115	1179	2964	1217	35	5	35	5	15	5	15	5

Russia (Pechora River)

Year	Declared	Declared	Return		Estimated %	Uncertainty in		Uncertainty in %	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW		estimate MSW	unreported	% unreported	unreported	unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
					salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	604.989	17728.011		NA	20	10	20	10	65	15	65	15
1972	825	24175		NA	20	10	20	10	65	15	65	15
1973	1705.011	49961.989 I		NA	20	10	20	10	65	15	65	15
1974	1320	38680		NA	20	10	20	10	65	15	65	15
1975	1298.352	38045.648		NA	20	10	20	10	65	15	65	15
1976	990.78	34394.22	NA	NA	20	10	20	10	65	15	65	15
1977	589.484	20463.516	NA	NA	20	10	20	10	65	15	65	15
1978	758.8	26341.2	AV	NA	20	10	20	10	65	15	65	15
1979	420.98	14614.02	AV	NA	20	10	20	10	65	15	65	15
1980	1123.472	39000.528	AV	NA	20	10	20	10	65	15	65	15
1981	126	20874	AV	NA	20	10	20	10	65	15	65	15
1982	54.4	13545.6 I	AV	NA	20	10	20	10	65	15	65	15
1983	597.6	16002.4	AV	NA	20	10	20	10	65	15	65	15
1984	1833.4	15966.6 I	AV	NA	20	10	20	10	65	15	65	15
1985	2762.5	29737.5	AV	NA	20	10	20	10	65	15	65	15
1986	65.6	32734.4	AV	NA	20	10	20	10	65	15	65	15
1987	21.2	21178.8	AV	NA	20	10	20	10	65	15	65	15
1988	3184	12816	NA.	NA	20	10	20	10	65	15	65	15
1989 N	A	NA	24596	27404	10	5	10	5	65	15	65	15
1990 N	A	NA	50	49950	10	5	10	5	65	15	65	15
1991 N	A	NA	7975	47025	10	5	10	5	65	15	65	15
1992 N	A	NA	550	54450	10	5	10	5	65	15	65	15
1993 N	A	NA	68	67932	10	5	10	5	65	15	65	15
1994 N	A	NA	3900	48100	10	5	10	5	65	15	65	15
1995 N	A	NA	9280	70720	10	5	10	5	65	15	65	15
1996 N		NA	8664	48336	10	5	10	5	65	15	65	15
1997 N		NA	1440	38560	10	5	10	5	65	15	65	15
1998 N		NA	780	59220	10	5	10	5	65	15	65	15
1999 N		NA	2120	37880	10	5	10	5	65	15	65	15
2000 N		NA	84	83916	10	5	10	5	65	15	65	15
2001 N		NA	2244	41756	10	5	10	- 5	65	15	65	15
2002 N		NA	405		10	5	10	-	65	15	65	15
2003 N		NA	1650		10	5	10	5	65	15	65	15
2004 N		NA	6075		10	5	10		65	15	65	15
2005 N		NA	2852	28148	10	5	10	5	65	15	65	15
2006 N		NA	1472		10	5	10	. 5	65	15	65	15
2007 N		NA	817	42183	10	5	10	5	65	15	65	15
2008 N		NA	300	49700	10	5	10	5	65	15	65	15
2009 N		NA	1116		10	5	10	5	65	15	65	15
2010 N		NA NA	1096		10	5	10	5	65	15	65	15
2010 N		NA NA	2990		10	3	10	5	65	15	65	15
2011 N		NA NA	4424	27176	10	3	10	5	65	15	65	15
2012 N		NA NA	4424	30983	10	5	10	5	65	15	65	15
2013 N 2014 N					10			5				15
	A	NA	2251	31349	10	5	10	5	65	15	65	15

Sweden

Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW		rate (%) - MSV
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmoi
1971	6220	254	30	15	30	15	52.5	12.5	57.5	12.
1972	4943	201	30	15	30	15	52.5	12.5	57.5	12.
1973	6124	895	30	15	30	15	52.5	12.5	57.5	12.
1974	8870	563	30	15	30	15	52.5	12.5	57.5	12.
1975	9620	160	30	15	30	15	52.5	12.5	57.5	12.
1976	5420	480	30	15	30	15	52.5	12.5	57.5	12.
1977	2453	206	30	15	30	15	52.5	12.5	57.5	12.5
1978	2903	254	30	15	30	15	52.5	12.5	57.5	12.5
1979	2988	661	30	15	30	15	52.5	12.5	57.5	12.5
1980	3842	1283	30	15	30	15	52.5	12.5	57.5	12.5
1981	7013	284	30	15	30	15	52.5	12.5	57.5	12.5
1982	6177	1381	30	15	30	15	52.5	12.5	57.5	12.5
1983	8222	903	30	15	30	15	52.5	12.5	57.5	12.5
1984	11584	1266	30	15	30	15	52.5	12.5	57.5	12.
1985	13810	470	30	15	30	15	52.5	12.5	57.5	12.
1986	14415	240	30	15	30	15	52.5	12.5	57.5	12.5
1987	11450	1084	30	15	30	15	52.5	12.5	57.5	12.5
1988	9604	1160	30	15	30	15	52,5	12.5	57.5	12.5
1989	2803	4044	30	15	30	15	52,5	12.5	57.5	12.5
1990	6839	2249	15	10	15	10	45	15	50	15
1991	8599	3033	15	10	15	10	45	15	50	15
1992	9550	4205	15	10	15	10	45	15	50	15
1993	9468	4762	15	10	15	10	45	15	50	15
1994	7347	3628	15	10	15	10	45	15	50	15
1995	8933	1528	15	10	15	10	37.5	12.5	42.5	12.5
1996	5318	2507	15	10	15	10	37.5	12.5	42.5	12.5
1997	2415	1809	15	10	15	10	37.5	12.5	42.5	12.5
1998	1953	1000	15	10	15	10	37.5	12.5	42.5	12.5
1999	3075	712	15	10	15	10	37.5	12.5	42.5	12.5
2000	5660	2546	15	10	15	10	37.5	12.5	42.5	12.5
2001	3504	3026	15	10	15	10	37.5	12.5	42.5	12.5
2002	3374	2075	15	10	15	10	37.5	12.5	42.5	12.5
2003	1833	496	15	10	15	10	37.5	12.5	42.5	12.
2004	1537	1528	15	10	15	10	37.5	12.5	42.5	12.
2005	1503	1027	15	10	15	10	37.5	12.5	42.5	12.
2006	1676	1069	15	10	15	10	37.5	12.5	42.5	12.
2007	521	1001	15	10	15	10	37.5	12.5	42.5	12.
2008	615	1112	12,5	7,5	12.5	7.5	27.5	12.5	32.5	12.
2009	651	979	12,5	7.5	12.5	7.5	27.5	12.5	32.5	12.
2010	1111	1139	12.5	7.5	12.5	7.5	27.5	12.5	32.5	12.
2011	1460	3100	17.5	7.5	17.5	7.5	45	15	50	1
2012	1336	3130	12.5	7.5	10	5	27.5	12.5	32.5	12.
2013	874	1431	10	5	10	5	30	15	35	15
2014	2347	2797	12.5	7.5	12.5	7.5	30	12.5	35	12.5
2015	1028	2569	12.5	7.5	12.5	7.5	30	12.5	30	12.5

UK (England and Wales)

Year	Declared	Estimated	Declared	Declared	Declared	Estimated	Estimated %	Uncertainty in %	Estimated %	Uncertainty	Estimated	Uncertainty in	Estimated	Uncertainty in	Estimated %	Estimated	Estimated	Estimate
	total catch	proportion	catch in NE	catch in NE			unreported	unreported	unreported catch	in %	exploitation		exploitation	exploitation	unreported catch	proportion	proportion	proportio
		1SW (total)	coastal	coastal	coastal	1SW (NE		catch of 1SW	of MSW salmon				rate (%) - MSW		in NE fishery			Scottish fish in N
				ishery - drift		fishery)	salmon	salmon		catch of MSW	1SW salmon	salmon	salmon	salmon				fishery (T/J nets
			total	nets	nets					salmon						(total)	(drift)	
1971		0.55048634	60353 N		NA	0.55048634	38.33057737	9.582644341	38.33057737		57.318052	10	42.49434582	10	32.27091633	0.95 N		NA
1972		0.41733812	51681 N		NA	0.41733812	38.96202931	9.740507327	38.96202931	9.740507327	51.31436424	10	37.84325293	10	32.27091633	0.95 N		NA
1973		0.52625241	62842 N		NA	0.52625241	38.3788506	9.59471265	38.3788506				37.25038117	10	32.27091633	0.95 N		NA
1974 1975		0.65362252 0.58546213	52756 N 53451 N		NA NA	0.65362252 0.58546213	39.30829568 38.5033576	9.827073919 9.6258394	39.30829568 38.5033576		49.79421655		36.99204217 36.71186529	10 10	32.27091633 32.27091633	0.95 N 0.95 N		NA NA
1976		0.64302775	15701 N		NA	0.64302775		9.187599446	36.75039778			10	37.13242587	10	32.27091633	0.95 P		NA
1977		0.61725763	52888 N		NA	0.61725763	39.02150715	9.755376788	39.02150715		50.37179751	10	37.23405997	10	32.27091633	0.93 N		NA
1978		0.68814567	51630 N		NA	0.68814567	38.44104806	9.610262015	38.44104806			10	36.16479365	10	32.27091633	0.92 N		NA
1979		0.81014211	43464 N		NA	0.81014211	38.55488888	9.638722219	38.55488888		47,72815082	10	35.1513067	10	32.27091633	0.91		NA
1980	90218	0.5507876	45780 N		NA	0.5507876	39.09573904	9.773934761	39.09573904			10		10	32.27091633	0.9 N		NA
1981		0.47731635	69113 N		NA	0.47731635	38.25011396	9.562528491	38.25011396		47.3632018	10	34.8634392	10	32.27091633	0.89 N		NA
1982	80289	0.66788214	50167 N		NA	0.66788214	38.27097871	9.567744678	38.27097871	9.567744678	47.26479216	10	34.82740717	10	32.27091633	0.88	۱A	NA
1983	116995	0.72145925	77277 N	IA	NA	0.72145925	37.13258923	9.283147308	37.13258923	9.283147308	47.14103398	10	34.65431849	10	32.27091633	0.87 N	IA	NA
1984	94271	0.73763119	59295 N	IA	NA	0.73763119	36.47361512	9.11840378	36.47361512	9.11840378	47.39983943	10	34.84849806	10	32.27091633	0.86	IA	NA
1985	95531	0.65753918	57356 N	IA	NA	0.65753918	38.91922644	9.729806609	38.91922644		47.50060204	10	34.86009626	10	32.27091633	0.85	IA.	NA
1986		0.62060623	63425 N		NA	0.62060623	37.97089883	9.492724706		9.492724706	46.8776795	10	34.30703129	10	32.27091633	0.84 N		NA
1987	83439	0.68	36143 N		NA	0.68	38.19270993	9.548177482	38.19270993		46.14426777	10	33.73304905	10	32.27091633	0.83 N		NA
1988	110163	0.69		47465	3384	0.69	39.671768	9.917942	39.671768	9.917942		10	33.51468908	10	32.27091633		0.82	0.
1989	83668	0.65		36236	5217	0.65	36.93679152	9.234197879	36.93679152			10	33.28875714	10	32.27091633		0.81	0.
1990	86676	0.52		48219	3311	0.52	36.6723293		36.6723293			10	33.24828948	10	31.34087237		0.8	0.
1991	51649	0.71		22463	2966	0.71	37.26469159	9.316172898	37.26469159		44.01914298	10	32.26829252	10	29.71559333		0.79	0.
1992 1993	44586 69177	0.77 0.81		17574 39224	2570 2576	0.77 0.81	39.82240193	9.955600481 9.490338166	39.82240193 37.96135267			10 10	31.76004191 29.49247154	10 10	28.02779616 26.27345845		0.78 0.77	0.
1993	88121	0.81		41298	5256		37.96135267 23.90506857	5.976267143	23.90506857			10	29.49247134	10	24.44821732		0.77	0.
1995	80478	0.77		48005	5205	0.77	22.34273059	5.585682647	22.34273059			10	27.1395055	10	22.54733219		0.76	0.
1996	46696	0.72		15172	3409	0.72	20.5945453	5.148636326	20.5945453		35.8495074	10	25.84879854	10	20.56564246		0.75	0.
1997	41374	0.73		19241	2681	0.73	18.83168136		18.83168136			10	23.90494707	10	18.49751949		0.75	0.
1998	36917	0.82		17328	937	0.82	18.92887505	4.732218764	18.92887505		31.36	10	22.4	10	18.49751949		0.75	0.
1999	41094	0.68		24812	2021	0.68	17.43967971	4.359919927	17.43967971			10	17.90300053	8.951500265	17.14697406		0.75	0
2000	60953	0.79	NA	40059	3295	0.79	14.88134535	3.720336338	14.88134535	3.720336338	29.68870689	10	14.99078601	7.495393005	13.11357936	NA	0.75	0.
2001	51307	0.75	NA	32374	3741	0.75	14.78866861	3.697167153	14.78866861	3.697167153	27.93381263	10	14.25662805	7.128314026	13.11357936	NA	0.75	0
2002	45669	0.76	NA	27685	3295	0.76	15.31083607	3.827709017	15.31083607	3.827709017	27.76935984	10	14.08814528	7.044072638	13.91678623	NA	0.75	0.
2003	22206	0.66	NA	5511	4924	0.66	17.43812388	4.35953097	17.43812388	4.35953097	21.43509617	10	10.67201679	5.336008395	17.14697406	NA	0.75	0.
2004	30559	0.81	NA	5921	5096	0.81	17.67995132	4.41998783	17.67995132	4.41998783	22.10898012	10	10.6330588	5.316529402	17.14697406	NA	0.75	0.
2005	26162	0.76	NA	5607	3380	0.76	17.56172605	4.390431513	17.56172605	4.390431513	21.82074844	10	10.58083843	5.290419214	17.14697406	NA	0.75	0
2006	22056	0.78		4040	3526	0.78	17.58807875	4.397019688	17.58807875				9.125043924	4.562521962	17.14697406		0.75	0
2007	19914	0.78		4894	2197	0.78		4.419821347	17.67928539				8.356540062	4.178270031	17.14697406		0.75	0
2008	19036	0.76		3649	2592	0.76	17.7675524	4.441888099	17.7675524			8.792928167	8.194308232	4.097154116	17.14697406		0.75	0
2009	13910	0.72		2590	2805	0.72	11.436047	2.85901175	11.436047		17.35706757	8.678533783	8.150014294	4.075007147	7.442387609		0.75	0
2010	32695	0.78		12214	7768	0.78	10.83235217	2.708088042	10.83235217				8.006827786	4.003413893	7.442387609		0.75	0
2011	34575	0.57		14915	9233	0.57	10.52969815	2.632424539	10.52969815			0 202221174	10.18721379	5.093606893	7.442387609		0.75	0
2012	14926	0.5		3571	3705	0.5		2.879247704	11.51699082				8.000087127	4.000043564	7.442387609		0.75	0
2013	22608	0.58		7964 6974	8679	0.58	9.852394246		9.852394246		17.35602817	8.678014083	8.526656606	4.263328303	7.442387609		0.75	0.
2014	14219 19732	0.54 0.47		9233	3826 6657	0.54 0.47	9.620814323 9.341568172	2.405203581 2.335392043	9.620814323	2.405203581 2.335392043	15.82894015	7.914470073 7.604719818	8.016755488 7.722899988	4.008377744 3.861449994	7.442387609 7.442387609		0.75 0.75	0.

UK (Northern Ireland)-Foyle Fisheries Area

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Year	Declared	Declared	Declared rod	Declared rod	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	net catch	net catch	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	1SW salmon	MSW	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
		salmon			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmor
1971	78037.23	5873.77 NA	A	NA	21.5	11.5	21.5	11.5	80	5	50	į
1972	64662.9	4867.1 NA	4	NA	21.5	11.5	21.5	11.5	80	5	50	
1973	57469.35	4325.65 NA	A	NA	21.5	11.5	21.5	11.5	80	5	50	5
1974	72587.43	5463.57 NA	A	NA	21.5	11.5	21.5	11.5	80	5	50	5
1975	51060.72	3843.28 NA	4	NA	21.5	11.5	21.5	11.5	80	5	50	5
1976	36205.83	2725.17 NA	A	NA	21.5	11.5	21.5	11.5	80	5	50	5
1977	36509.94	2748.06 NA	A	NA	21.5	11.5	21.5	11.5	80	5	50	5
1978	44557.23	3353.77 NA	4	NA	21.5	11.5	21.5	11.5	80	5	50	5
1979	34412.79	2590.21 NA	A	NA	21.5	11.5	21.5	11.5	80	5	50	5
1980	45777.39	3445.61 NA	A	NA	21.5	11.5	21.5	11.5	80	5	50	5
1981	32346.33	2434.67 NA	A	NA	21.5	11.5	21.5	11.5	80	5	50	5
1982	55946.01	4210.99 NA	A	NA	21.5	11.5	21.5	11.5	80	5	50	5
1983	77424.36	5827.64 NA		NA	21.5	11.5	21.5	11.5	80	5		5
1984	27464.76	2067.24 NA		NA	21.5	11.5	21.5	11.5	80	5	50	5
1985	37684.53	2836.47 NA		NA	21.5	11.5	21.5	11.5	80	5		5
1986	43109.22	3244.78 NA		NA	21.5	11.5	21.5	11.5	80	5		5
1987	17189.19	1293.81 NA		NA	21.5	11.5	21.5	11.5	69	7		5
1988	43974.12	3309.88 NA		NA	21.5	11.5	21.5	11.5	64.5	6.5		4
1989	60288.18	4537.82 NA		NA	23.5	13.5	23.5	13.5	89	9		6
1990	39874.68	3001.32 NA		NA	13.5	3.5	13.5	3.5	62	6		4
1991	21708.99	1634.01 NA		NA	13.5	3.5	13.5	3.5	64.5	6.5		4
1992	39299.01	2957.99 NA		NA	16.5	6.5	16.5	6.5	56			3
1993	35366.04	2661.96 NA		NA	13.5	3.5	13.5	3.5	41			1
1994	36143.52	2720.48 NA		NA	19.5	9	19	3.3	70			
1995	33398.16	2513.84 NA		NA	13.5	3.5	13.5	3.5	67	7		- 4
1996	28405.92	2138.08 NA		NA	15.5	5.5	15.5	5.3		10		10
1997	40885.59	3077.41 NA		NA	10	5	10	5	60			10
1998	37153.5	2796.5 NA		NA	10	5	10	5	25	5		7.5
1999	21659.7	1630.3 NA		NA	10	5	10	5	63	5		7.5
2000	30384.96	2287.04 NA		NA NA	10	5	10	5	58			7.5
					5	5	5					
2001	21367.68 37914.24	1608.32 NA		NA 690	2.5	2.5		2.5	50 15			5
2002	30440.76	2853.76	9163		0.5	0.5	2,5 0,5	0.5				3
		2291.24	4576				0.5	$\overline{}$	15	3		
2004	20729.7	1560.3	4570		0.5	0.5	_	0.5				3
2005	23746	1787	7079		0.5	0.5	0.5	0.5	15			3
2006	11324	852	4886		0.5	0.5	0.5	0.5	15			3
2007	5050	322	9530		0.5	0.5	0.5	0.5	15			- 3
2008	3880	292	4755		0.5	0.5	0.5	0.5	15			3
2009	1743	194	3640		0.5	0.5	0.5	0.5	15			3
2010	0	0	4257		0.5	0.5	0.5	0.5	15			3
2011	0	0	3770		1		1	1	15			
2012		0	4781		1	1	1	1	10			7.5
2013	0	0	2831	283	1	1	1	1	10			7.5
2014	0	0	1436		1	1	1	1	10			7.5
2015	0	0	3241	360	1	. 1	1	1	10	7.5	10	7.5



UK (Northern Ireland)-DCAL area

rear	Declared	Declared	Declared rod	Declared rod	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in	Fish P	roportion 1SW:	Fish	Proportion 1SW:	Scaling factor for	Scaling factor	Scaling factor for	Scaling factor
	net catch	net catch	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation	counter;	river Bush	counter:	river Bann	1SW estimate	error for 1SW	MSW estimate	error for MSW
	1SW salmon	MSW	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW	river Bush		river Bann			estimate		estimate
		salmon			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon								
L971	35506	2673 N		NA	21.5	11.5	21.5	11.5	80	5	50	5		0						NA
1972	34550	2601 N		NA	21.5	11.5	21.5	11.5	80	5	50	5	0	0	0		NA		NA	NA
.973	29229	2200 N		NA	21.5	11.5	21.5	11.5	80	5	50	5	0	0	0		NA		NA	NA
974	22307	1679 N		NA	21.5	11.5	21.5	11.5	80	5	50	3	0	0			NA			NA
975	26701	2010 N		NA	21.5	11.5	21.5	11.5	80	5	50	5	0	0	-		NA		NA	NA
976	17886	1346 N		NA	21.5	11.5	21.5	11.5	80	5	50	5		0	-		NA		NA	NA
977	16778	1263 N		NA	21.5	11.5	21.5	11.5	80	5	50	5	4	0	-		NA		NA	NA
978	24857	1871 N		NA	21.5	11.5	21.5	11.5	80	5	50	5		0	-		NA		NA	NA
979	14323	1078 N		NA	21.5	11.5	21.5	11.5	80	5	50	5		0	-		NA		NA	NA
980	15967	1202 N		NA	21.5	11.5	21.5	11.5	80	5	50	5		0	-		NA		NA	NA
981 982	15994 14068	1204 N		NA	21.5 21.5	11.5 11.5	21.5	11.5 11.5	80 80	5	50	- 5	0	0			NA NA		NA	NA NA
	20845	1059 N		NA	21.5	11.5	21.5 21.5	11.5	80	5	50	5	-	0	-				NA NA	NA NA
983		1569 N		NA							50	5					NA			
984 985	11109	836 N		NA	21.5	11.5	21.5	11.5	80	5	50	5	-	0	-		NA		NA	NA
985	12369 13160	931 N		NA NA	21.5 21.5	11.5	21.5	11.5 11.5	80 80	5	50 50	5	-	0	-		NA NA		NA NA	NA NA
985	9240	991 N 695 N		NA NA	21.5	11.5 11.5	21.5 21.5	11.5	69	7	50	5	-	83	-					
988	14320	1078 N		NA NA	21.5	11.5	21.5	11.5	64.5		36	4		96		0			0.0	
989	15081	1135 N		NA	23.5	13.5	23.5	13.5	90	0.5	60	6		82		0		_	0.0	
990	9499	715 N		NA NA	13.5	3.5	13.5	3.5	62		38	4		87		0				
991	6987	526 N		NA	13.5	3.5	13.5	3.5			43	4		87		0				
992	9346	703 N		NA	16.5	6.5	16.5	6.5		6.5	33	3		84		0				
993	7906	703 N 595 N		NA NA	13.5	3.5	13.5	0.5	41	4	12	1		93		0			0.0	
994	11206	843 N		NA	19	9.5	19.3	9.5	70	7	40	4		88		0				
995	11637	876 N		NA	13.5	3.5	13.5	3.5	67	7	40	4		92		0	- 10			
996	10383	781 N		NA	15.5	5.5	15.5	5.5	57	10	34	10		87		0				
997	10479	789 N		NA	10	5	10	5	60	10	34	10		85		85				
998	9375	706 N		NA	10	5	10		25	5	22.5	7.5		95		95				
999	9011	678 N		NA	10	5	10	5	63	5	32.5	7.5		90		90				
000	10598	798 N		NA	10	5	10	5	58	5	32.5	7.5		91		91				
001	8104	610 N		NA	5	5	5	5		5	30	5		97		97				
002	3315	249	2218		2.5	2.5	2.5	2.5		8.75	13.7264853	8.75		95		95			61	
003	2236	168	1884		2.5	2.5	2.5	2.5		6.6	12.26040087	6.6		96		96				
004	2411	181	3053	230	0.5	0.5	0.5	0.5	18.31195792	9.7	18.31195792	9.7		92		92	67	5	61	
005	3012	227	1791	135	0.5	0.5	0.5	0.5	11.89297533	7.05	11.89297533	7.05	1151	91		91		5	61	
06	2288	172	1289		0.5	0.5	0.5	0.5	12.43353682	7.95	12.43353682	7.95	1074	87		87		5		
007	2533	162	2427	155	0.5	0.5	0.5	0.5	11.03249511	3.55	11.03249511	3.55	2584	94	7008	94	67	5	61	
108	1825	116	2444		0.5	0.5	0.5	0.5	13.93543107	7.1	13.93543107	7.1	1712	90		0	10	2	9.5	
109	1383	154	1457	162	0.5	0.5	0.5	0.5	9.85798445	2.95	9.85798445	2.95		83	3838	83	67	5	61	
10	1723	191	1327		0.5	0.5	0.5	0.5		2.5	14.60830975	2.5		78		70	67	5	61	
11	857	285	1132		1	1	1	1	15	5	15	5		73		76		5		
112	15	5	263		1	1	1	1	10	5	10	5		74		71		5	61	
13	9	1	46	5	1	1	1	1	0.625	0.375	0.625	0.375	1644	92	5866	91	67	5	61	
14	0	0	143	40	2.5	2.5	2.5	2.5	0.625	0.375	0.625	0.375	963	76	4335	91	67	5	61	
15	0	0	20	6	2.5	2.5	2.5	2.5	0.625	0.375	0.625	0.375	1005	83	6235	86	67	5	61	

UK (Scotland)-East

ionic c	arlo simulatio	, <u>,</u>								
Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	216873	135530	25	10	25	10	75.38	12.56	49.89	9.98
1972	220106	183875	25	10	25	10	76.82	12.8	51.45	10.29
1973	259773	204826	25	10	25	10	74.91	12.49	49.87	9.97
1974	245424	158959	25	10	25	10	81.97	13.66	56.35	11.2
1975	181940	180828	25	10	25	10	80.47	13.41	55.06	11.0
1976	150069	92179	25	10	25	10	76.55	12.76	50.66	10.13
1977	154306	118645	25	10	25	10	81.44	13.57	55.78	11.16
1978	158859	139763	25	10	25	10	75.57	12.6	50.97	10.19
1979	160796	116559	25	10	25	10	78.36	13.06	53.87	10.77
1980	101665	155646	17.5	7.5	17.5	7.5	76.76	12.79	51.96	10.39
1981	129690	156683	17.5	7.5	17.5	7.5	75.9	12.65	51.19	10.24
1982	175374	113198	17.5	7.5	17.5	7.5	71.07	11.84	45.26	9.05
1983	170843	126104	17.5	7.5	17.5	7.5	77	12.83	49.38	
1984	175675	90829	17.5	7.5	17.5	7.5	70.12	11.69	43.89	8.78
1985	133119	95044	17.5	7.5	17.5	7.5	61.86	10.31	38.93	7.79
1986	180292	128654	17.5	7.5	17.5	7.5	59.49	9.92	37.55	7.53
1987	139259	88531	17.5	7.5	17.5	7.5	64.52	10.75	40.51	8.3
1988	118620	91167	17.5	7.5	17.5	7.5	40.3	6.72	29.19	5.84
1989	143063	85399	10	5	10	5	37.54	6.26	27.96	
1990	63352	73974	10	5	10	5	39.84	6,64	28.74	5.75
1991	53861	53693	10	5	10	5	36.78	6.13	27.45	5.49
1992	79906	67979	10	5	10	5	32.12	5.35	25.86	5.1
1993	73413	60498	10	5	10	5	35.33	5.89	26.93	5.39
1994	80429	72758	10	5	10	5	33.08	5.51	26.07	5.2
1995	73020	69055	10	5	10	5	30.9		25.39	5.08
1996	56627	50365	10	5	10	5	28.78	4.8	24.47	4.89
1997	37448	34850	10	5	10	5	30.58	5.1	25.14	5.03
1998	44952	32231	10	5	10		24.23	4.04	22.92	4.58
1999	20907	27011	10	5	10	5	24.8	4.13	23.33	4.67
2000	36871	31280	10	5	10	5	21.85	3.64	22.26	
2001	36646	30470	10	5	10	5	20.44	3.41	21.59	4.51
2002	26616	21740	10	5	10	5	19.27	3.21	21.16	
2003	25871	24270	10	5	10		17.25	2.75	19.25	4.25
2003	31667	30773	10	5	10	5	17.25	2.75	19.25	4.25
2005	31597	23676	10	. 5	10	5	17.25	2.75	19.25	4.25
2006	30739	22954	10	5	10	5	15.25	2.75	16.5	3.5
2007	26015	19444	10	_ 5	10	5	13.75	2.75	15	3.5
2008	18586	20757	10	5	10	5	10.75	2.75	14	3.5
2009	14863	15042	10	5	10	5	9.75	2.75	13	3.5
2010	28161	22861	10	5	10	5	9.75	2.75	13	3.5
2010	12485	24213	10	5	10	5	9.75	2.75	12.5	3.5
2011	16117	16165	10	5	10	5	8.25	2.75	11.5	3.5
						5				
2013	18400	14901	10	5 5	10	5	7.25	2.75	11	3.5
2014 2015	10922 12153	11795 7734	10 10		10 10	5	5.75 5.75	2.25 2.25	9.5 9.5	3

UK (Scotland)-West

Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	45287	26071	35	10	35	10	37.69	6.28	24.94	4.99
1972	31358	34148	35	10	35	10	38.41	6.4	25.72	5.14
1973	33317	33094	35	10	35	10	37.46	6.24	24.94	4.99
1974	43992	29369	35	10	35	10	40.99	6.83	28.17	5.63
1975	40424	27145	35	10	35	10	40.24	6.71	27.53	5.51
1976	38409	22367	35	10	35	10	38.27	6.38	25.33	5.07
1977	39952	20335	35	10	35	10	40.72	6.79	27.89	5.58
1978	45611	23191	35	10	35	10	37.79	6.3	25.49	5.1
1979	26440	15950	35	10	35	10	39.18	6.53	26.94	5.39
1980	19776	16942	27.5	7.5	27.5	7.5	38.38	6.4	25.98	5.2
1981	21033	18021	27.5	7.5	27.5	7.5	37.95	6.33	25.6	5.12
1982	32687	15044	27.5	7.5	27.5	7.5	35.53	5.92	22.63	4.53
1983	38774	19857	27.5	7.5	27.5	7.5	38.5	6.42	24.69	4.94
1984	37404	16384	27.5	7.5	27.5	7.5	35.06	5.84	21.94	4.39
1985	24861	19571	27.5	7.5	27.5	7.5	30.93	5.15	19.46	3.89
1986	22546	19543	27.5	7.5	27.5	7.5	29.75	4.96	18.78	3.76
1987	25506	15460	27.5	7.5	27.5	7.5	32.26	5.38	20.26	4.05
1988	30463	20989	27.5	7.5	27.5	7.5	20.15	3.36	14.6	2.92
1989	31878	18487	20	5	20	5	18.77	3.13	13.98	2.8
1990	17742	13950	20	5	20	5	19.92	3.32	14.37	2.87
1991	19747	11500	20	5	20	5	18.39	3.07	13.72	2.74
1992	21770	14862	20	5	20	5	16.06	2.68	12.93	2.59
1993	21104	11228	20	5	20	5	17.67	2.94	13.47	2.69
1994	18234	12304	20	5	20	5	16.54	2.76	13.03	2.61
1995	16784	9133	20	5	20	5	15.45	2.58	12.69	2.54
1996	9537	7463	20	5	20	5	14.39	2.4	12.24	2.45
1997	9059	5504	20	5	20	5	15.29	2.55	12.57	2.51
1998	8369	6150	20	5	20	5	12.12	2.02	11.46	2.29
1999	4147	3588	20	5	20	5	12.4	2.07	11.66	2.33
2000	6974	5301	20	5	20	5	10.92	1.82	11.13	2.23
2001	5603	4191	20	5	20	5	10.22	1.7	10.8	2.26
2002	4691	4548	20	5	20	- 5	9.64	1.61	10.58	2.12
2003	3536	3061	20	5	20	5	4.75	0.75	5.25	1.25
2004	5836	6024	20	5	20	5	7	1	7.5	1.5
2005	7428	4913	20	5	20	5	7	1	7.5	1.5
2006	5767	4403	20	5	20	5	7	1	7.5	1.5
2007	6178	4468	20	5	20	5	7	1	7.5	1.5
2008	4740	4853	20	5	20	5	7	1	7.5	1.5
2009	3250	4095	20	5	20	5	6	1	6.5	1.5
2010	5107	4051	20	5	20	5	6	1	6.5	1.5
2011	3206	4246	20	5	20	5	5.5	1	6	1.5
2012	3239	3392	20	5	20	5	4.5	1	5	1.5
2013	2342	2286	20	5	20	5	4	1	4.75	1.5
2014	1653	1490	20	5	20	5	3.5	1	4.5	1.5
2015	1423	1269	20	5	20	5	3.5	1	4.5	1.5

Faroes

araes aeri	ne uniform dist					
Year	Catch 1SW salmon	Catch MSW salmon	Estimated % unreported catch of 1SW	Uncertainty in % unreported catch of 1SW	% wild	Natural mortality aft 1 st sea winte
			salmon	salmon		(M)
1971	2,620	105,796	10.0	5.0	1.0	0.
1972	2,754	111,187	10.0	5.0	1.0	0.
1973	3,121	126,012	10.0	5.0	1.0	0.
1974	2,186	88,276	10.0	5.0	1.0	0.
1975	2,798	112,984	10.0	5.0	1.0	0
1976	1,830	73,900	10.0	5.0	1.0	0
1977	1,291	52,112	10.0	5.0	1.0	0
1978	974	39,309	10.0	5.0	1.0	0
1979	1,736	70,082	10.0	5.0	1.0	0
1980	4,523	182,616	10.0	5.0	1.0	0
1981	7,443	300,542	10.0	5.0	1.0	0
1982	6,859	276,957	10.0	5.0	1.0	0
1983	15,861	215,349	10.0	5.0	1.0	0
1984	5,534	138,227	10.0	5.0	1.0	0
1985	378	158,103	10.0	5.0	0.9	0
1986	1,979	180,934	10.0	5.0	1.0	0
1987	90	166,244	10.0	5.0	1.0	0
1988	8,637	87,629	10.0	5.0	0.9	0
1989	1,788	121,965	10.0	5.0	0.8	0
1990	1,989	140,054	10.0	5.0	0.5	0
1991	943	84,935	10.0	5.0	0.5	0
1992	68	35,700	10.0	5.0	0.6	0
1993	6	30,023	10.0	5.0	0.7	0
1994	15	31,672	10.0	5.0	0.7	0
1995	18	34,662	10.0	5.0	0.8	0
1996	101	28,381	10.0	5.0	0.8	C
1997	101	20,381	0.0	0.0	0.8	0
1998	339	1,424	15.0	5.0	0.8	0
1999	333	- 1,424	0.0	0.0	0.8	C
2000	225	1,765	15.0	5.0	0.8	0
2000	-	- 1,705	0.0		0.8	0
				0.0		
2002	-	-	0.0	0.0	0.8	0
2003	-	-	0.0	0.0	0.8	0
2004	-	-	0.0	0.0	0.8	0
2005	-	-	0.0	0.0	0.8	0
2006	-	-	0.0	0.0	0.8	0
2007	-	-	0.0	0.0	0.8	0
2008	-	-	0.0	0.0	0.8	0
2009	-	-	0.0	0.0	0.8	0
2010	-	-	0.0	0.0	0.8	0
2011	-	-	0.0	0.0	0.8	0
2012	-	-	0.0	0.0	0.8	0
2013	-	-	0.0	0.0	0.8	0
2014	-	-	0.0	0.0	0.8	0
2015		_	0.0	0.0	0.8	0

West Greenland

Year	Declared	Estimated	Wean weight	Estimated	Estimated	Proportion	Proportion	No. Fish	No. Fish
	catch (t)	unreported		min'	max'	1SW in NAC	1SW in NEAC	identified as	identified as
		catch		proportion of	proportion of	fish	fish	NAC (from	NEAC (from
				NAC fish	NAC fish			genetic	genetic
				(from scale	(from scale			analysis)	analysis)
				analysis)	analysis)				
971	2689	0	3.14	0.28	0.40	0.945	0.964	-	-
972	2113	0	3.44	0.34	0.37	0.945	0.964	-	-
973	2341	0	4.18	0.39	0.59	0.945	0.964	-	-
974	1917	0	3.58	0.39	0.46	0.945	0.964	-	-
975	2030	0	3.12	0.40	0.48	0.945	0.964	-	-
976	1175	0	3.04	0.38	0.48	0.945	0.964	-	-
977	1420	0	3.21	0.38	0.57	0.945	0.964	-	
978	984	0	3.35	0.47	0.57	0.945	0.964	-	-
979	1395	0	3.34	0.48	0.52	0.945	0.964	-	-
980	1194	0	3.22	0.45	0.51	0.945	0.964	-	-
981	1264	0	3.17	0.58	0.61	0.945	0.964	-	-
982	1077	0	3.11	0.60	0.64	0.945	0.964	-	-
983	310	0	3.10	0.38	0.41	0.945	0.964	-	-
984	297	0	3.11	0.47	0.53	0.945	0.964	-	-
985	864	0	2.87	0.46	0.53	0.925	0.950	·	-
986	960	0	3.03	0.48	0.66	0.951	0.975		-
987	966	0	3.16	0.54	0.63	0.963	0.980	-	-
988	893	0	3.18	0.38	0.49	0.967	0.981	-	-
989	337	0	2.87	0.52	0.60	0.923	0.955		-
990	274	0	2.69	0.70	0.79	0.957	0.963	-	-
991	472	0	2.65	0.61	0.69	0.956	0.934		-
992	237	0	2.81	0.50	0.57	0.919	0,975	X	-
993	0	12	2.73	0.50	0.76	0.95	0.96		-
994	0	12	2.73	0.50	0.76	0.95	0.96	-	-
995	83	20	2.56	0.65	0.72	0.968	0.973	-	-
996	92	20	2.88	0.71	0.76	0.941	0.961	-	-
997	58	5	2.71	0.75	0.84	0.982	0.993	-	-
998	11	11	2.78	0.73	0.84	0.968	0.994	-	-
999	19	12.5	3.08	0.84	0.97	0.968	1.000	-	-
2000	21	10	2.57	0	0	0.974	1.000	344	146
2001	43	10	3.00	0.67	0.71	0.982	0.978	1	1
2002	9.8	10	2.90	0	0	0.973	1.000	338	163
2003	12.3	10	3.04	_ 0	0	0.967	0.989	1212	567
2004	17.2	10	3.18	0	0	0.970	0.970	1192	447
2005	17.3	10	3.31	0	0	0.924	0.967	585	182
2006	23.0	10	3.24	0	0	0.930	0.988	857	326
2007	24.8	10	2.98	0	0	0.965	0.956	917	206
2008	28.6	10	3.08	0		0.974	0.988	1593	260
2009	28.0	10	3.50	0	0	0.934	0.894	1483	138
2010	43.1	10	3.42	0	0	0.982	0.975	991	249
2011	27.4	10	3.40	0	0	0.939	0.831	888	72
012	34.6	10	3,44	0.00	1.00	0.932	0.980	1121	252
2013	47.7	10	3.35	0.00	1.00	0.949	0.966	938	211
2014	70.4	10	3.32	0.00	1.00	0.913	0.961	660	260
015	60.9	10	3.37	0.00	1.00	0.970	0.982	1337	337

Stock composition					
Country	MSW				
France	0.027				
Finland	0.001				
Iceland	0.001				
Ireland	0.147				
Norway	0.028				
Russia	0.000				
Sweeden	0.003				
UK(England & Wales)	0.149				
UK(Northern Ireland)	0.000				
UK(Scotland)	0.644				
Other					
Total	1.00				

Appendix 4: Input data for Atlantic salmon used to do the runreconstruction and estimates of returns and spawners by size group and age group for North America

Appendix 4.i. Input data for the fishery at West Greenland used in the run reconstruction model.

	Harvest in tonnes of salmon at West	Unreported harvest in	Mean weight of	Prop. of harvest of	Prop. of harvest of			Number of genetic	Number of genetic
Year of the	Greenland for	tonnes at West		NAC origin (min	NAC origin (max	Prop. of NAC origin	Prop. of NEAC origin	samples that are	samples that are
fishery	assessment	Greenland	harvest	based on scales)	based on scales)	salmon that are 1SW	salmon that are 1SW	NAC origin	NEAC origin
Year	WG.harv	WG.unhary	WG meanwt		WG.propNAC max		WG.prop1SWNEAC	WG.sampleNAC	WG.sampleNEAC
1971	2689				0.40	0.945			
1972	2113				0.37	0.945	0.964		
1973	2341	0			0.59		0.964		
1974	1917	0			0.46		0.964		
1975	2030				0.48		0.964		
1976	1175				0.48		0.964		
1977	1420			0.38	0.57	0.945	0.964		
1978	984	0			0.57	0.945	0.964		
1979	1395				0.52		0.964		
1980	1194				0.51	0.945	0.964		
1981	1264	0			0.61	0.945	0.964		
1982	1077	0		0.60	0.64	0.945	0.964		
1983	310				0.41	0.945	0.964		
1984	297	0		0.47	0.53		0.964		
1985	864	0			0.53		0.950		
1986	960				0.66		0.975		
1987	966				0.63		0.980		
1988	893				0.49		0.981		
1989	337	0			0.49		0.955		
1990	274				0.79		0.963		
1991	472				0.69				
1992	237	0		0.50	0.57	0.919			
1993	0				0.76				
1994	0				0.76		0.96		
1995	83				0.70		0.973		
1996	92				0.76		0.961		
1997	58			0.71	0.84		0.993		
1998	11				0.84		0.994		
1999	19				0.97		1.000		
2000	21	10		0.04	0.57				
2001	43				0.71	0.982			
2002	9.8				0.71				
2003	12.3				0		0.989		
2004	17.2				0		0.970		
2005	17.3			0			0.967		
2006	23.0						0.988		
2007	24.8				0				
2007	28.6				0				
2009	28.0				0				
2010	43.1	10			0		0.894		
2010	27.4				0			888	
2012	34.6				0				
2012	47.7	10			0		0.966		
2013	58.2				0				
2014	55.9			0	0		0.981		

Appendix 4.ii. Input data for sea fisheries on large salmon and small salmon from Newfoundland and Labrador used in the run reconstruction model. Labrador represents harvests from Labrador in aboriginal fisheries for food, social and ceremonial purposes and the resident food fishery beginning in 1998.

	Harvests of la	arge salmo	n (number)	Harvests of s	mall salmon (number)
		Salmon				
	Salmon	fishing	Labrador	Salmon	Salmon	Labrador
	fishing	areas 8 to	subsistence	fishing	fishing areas	subsistence
	areas 1 to 7	14a	fisheries	areas 1 to 7	8 to 14a	fisheries
	All - I D IN	NII - NIEGE		Name I Daniel	Name NEOF 4	
	Nlg_LBandN	<u> </u>	NI IDECO	_	Nsm_NF8to1	
/ear	F1to7	o14a	NIg_LBFSC	NF1to7	4a	Nsm_LBFSC
1970		-	0		0	
1971			0			
1972			0		111141	
1973			0		176907	
1974	196726	85714	0		153278	
1975	215025	72814	0	302348	91935	
1976	210858	95714	0	221766	118779	
1977	231393	63449	0	220093	57472	
1978	155546	37653	0	102403	38180	
1979	82174	29122	0	186558	62622	
1980	211896	54307	0	290127	94291	
1981	211006	38663	0	288902	60668	
1982	129319	35055	0	222894	77017	
1983	108430	28215	0	166033	55683	
1984	87742	15135	0	123774	52813	
1985	70970	24383	0	178719	79275	
1986	107561	22036	0	222671	91912	
1987	146242	19241	0	281762	82401	
1988	86047	14763	0	198484	74620	
1989			0		60884	
1990			0	104788	46053	
1991	39257	10259	0		42721	
1992			0		0	
1993			0		0	
1994			0		0	
1995			0		0	
1996		0	0		0	
1997			0		0	
1998			2269			
1999			1084			
2000			1352		0	
2000	0		1721		0	
2002	0		1389		0	
2003			2175		0	
2004			3696			
2005			2817			
2006			3090			
2007			2652			
2008			3909			
2009			3344			
2010			3725			
2011			4451			
2012			4228			
2013		0	6479		0	
2014	0	0	3994	0	0	895
2015	0	0	6146	0	0	892

Appendix 4.iii. Input data for sea fisheries on large salmon and small salmon from St Pierre & Miquelon used in the run-reconstruction model.

Number of sma		Number of	Reported harvest	Year of the
salmo	large salmon	salmon	(kg)	fishery
	0	0	0	1970
	0	0		1971
	0	0	0	1972
	0	0	0	1973
	0	0	0	1974
	0	0	0	1975
99	333	1331	3000	1976
	0	0	0	1977
	0	0	0	1978
	0	0	0	1979
	0	0	0	1980
	0	0	0	1981
	0	0	0	1982
99	333	1331	3000	1983
99	333	1331	3000	1984
99	333	1331	3000	1985
		1109	2500	
83	277		2000	1986
66	222	887		1987
66	222	887	2000	1988
66	222	887	2000	1989
63	211	843	1900	1990
39	133	532	1200	1991
76	255	1020	2300	1992
96	322	1287	2900	1993
113	377	1508	3400	1994
26	89	355	800	1995
53	177	710	1600	1996
49	166	665	1500	1997
76	255	1020	2300	1998
77	258	1030	2322	1999
75	251	1006	22,67	2000
71	239	956	2155	2001
65	217	866	1952	2002
			2892	
96	321	1283		2003
92	309	1235	2784	2004
109	365	1458	3287	2005
118	394	1577	3555	2006
64	216	864	1947	2007
117 115	393 384	1571 1535	3540 3460	2008 2009
92	308	1233	2780	2009
125	417	1667	3757	2010
48	161	643	1450	2012
176	588	2351	5300	2013
126	423	1690	3810	2014
105	500	1557	3510	2015

Appendix 4.iv. Input data for large salmon for Labrador used in the run-reconstruction.

	Large Salmon																	
	Commercial ha	arvest (number	r of fish)	FSC Fishery	Proportion Labr	rador origin		-			Exploitation r	ate	Proportion 2S	W	Returns to Labra	ador rivers	Angling catch	nes
	SFA 1	SFA 2	SFA 14B	ALL	SFA 1		SFA 2		SFA 14B		All SFAs				number of fish		number of fis	h
Year				Number of fish	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Retained	Release
1970	25127	64806	13673		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(562	
1971	21599	55708	11753		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(486	
1972	30204	77902	16436		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(424	
1973	13866	93036	15863		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(1009	
1974	28601	71168	14752		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(803	
1975	38555	77796	15189		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(327	
1976	28158	70158	18664		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(830	
1977	30824	48934	11715		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(1286	
1978	21291	27073	3874		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	C	(767	
1979	28750	87067	9138		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(609	
1980	36147	68581	7606		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(889	
1981	24192	53085	5966		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	O	(520	
1982	19403	33320	7489		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0		621	
1983	11726	25258	6218		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	C	(428	
1984	13252	16789	3954		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	C	(510	
1985	19152	34071	5342		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	C	(294	
1986	18257	49799	11114		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(467	
1987	12621	32386	4591		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	O	(633	
1988	16261	26836	4646		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	C)	710	
1989	7313	17316	2858		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	C	(461	
1990	1369	7679	4417		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	(357	
1991	9981	19608	2752		0.6	0.8	0:6	0.8	0.6	0.8	0.580	0.830	0.70	0.90	0	(93	
1992	3825	9651	3620		0.6	0.8	0.6	0.8	0.6	0.8	0.38	0.62	0.70	0.90	0	(781	1
1993	3464	11056	857		0.6	0.8	0,6	0.8	0.6	0.8	0.29	0.50	0.70	0.90	0	(378	
1994	2150	8714	312		0.6	0.8	0.6	0.8	0.6	0.8	0.14	0.25	0.70	0.90	0	(455	34
1995	1375	5479	418		0.6	0.8	0.6	0.8	0.6	0.8	0.13	0.23	0.70	0.90	0	(408	50
1996	1393	5550	263		0.6433	0.7247	0.8839	0.9521	0.6	0.8	0.17	0.30	0.70	0.90	0	(334	48
1997	6943		***************************************		1	1	1	1	1	1	0.17	0.30	0.60	0.71	0	(158	
1998	0	0	0	2269	1	1	1	1	1	1	0.17	0.30	0.60	0.71	7374	19486	231	
1999	0	0	0	1084	1		1	1	1	1	0.17	0.30	0.60	0.71	8827	23328	320	93
2000	0	0	0	1352	1	1	1	1	1	1	0.17	}	0.60	0.71	12052		~	}~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~
2001	0	0	0	1721	1	1	1	1	1	1	0.17	·	0.60	0.71	12744	<u> </u>	7 338	
2002	0	0	0	1389	1	$-\sqrt{1}$	1	1	1	1	0.17	·	0.60	0.71	9076	<u> </u>		g
2003	0	0	0	2175	1	<u></u>	1	1	1	1	0.17	(0.60	0.71	6676			§ and a second s
2004	0	0	0	3696	1	1	1	1	1	1	0.17	>	0.60	0.71	10964			}
2005	0	0	0	2817	1	1	1	1	1	1	0.17	<u> </u>	0.60	0.71	11159	}		129
2006	0	0	0	3090	1	1	1	1	1	1	0.17	} ~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~	0.60	0.71	12414			<u> </u>
2007	0	0	0	2652	1	1	1	1	1	1	0.17		0.60	0.71	11887	<u> </u>	3 235	§~~~~~
2008	0	0	0	3909	1	1	1	1	1	1	0.17	}	0.60	0.70	14700	-	7 200	}
2009	0	0	0	3344	1	1	1	1	1	1	0.2		0.60	0.70	18643			(
2010	0	0	0	3725	1	1	1	1	1	1	0.2	}	0.60	0.70	10764		~~~~~~~~~~~~~	}~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~
2011	0	0	0	4451	1	1	1	1	1	1	0.2	ţ	0.60	0.70	30198			
2012	0	0	0	4228	1	1	1	1	1	1	0.2	·	0.60	0.70	19062	<i></i>	0	g
2013	0	0	0	6479	1	1	1	1	1	1	0.2	0.4	0.60	0.70	36859	91394	~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~	§
2014	0	0	0	3994	1	1	1	1	1	1	0.2	0.4	0.60	0.70	36055	87989	0	<u> </u>
2015	0	0	0	6146	1	1	1	1	1	1	0.2	0.4	0.60	0.70	49662	127898	3 0	102

Appendix 4.v. Input data for small salmon for Labrador used in the run-reconstruction.

	Small salmor	n														
	Commercial h	harvest (numbe	r of fish)	FSC Fishery	Proportion La	brador origin					Exploitation ra	te	Returns to Labrac	for rivers	Angling catche	es
	SFA 1	SFA 2	SFA 14B	ALL	SFA 1		SFA 2		SFA 14B		All SFAs		number of fish		number of fish	***************************************
Year				Number of fish	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Retained	Released
1970	19109	38359	11212		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	4013	0
1971	14303	28711	8392		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	3934	0
1972	3130	6282	1836		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	2947	0
1973	9848	37145	9328		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	7492	0
1974	34937	57560	19294		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	2501	0
1975	17589	47468	13152		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	3972	0
1976	17796	40539	11267		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	5726	0
1977	17095	12535	4026		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	4594	0
1978	9712	28808	7194		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	2691	0
1979	22501	72485	8493		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	4118	0
1980	21596	86426	6658		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	3800	0
1981	18478	53592	7379		0.6	0.8	0.6	0,8	0.6	0.8	0.3	0.5	0	C	5191	0
1982	15964	30185	3292		0.6	0.8	0.6	0,8	0.6	0.8	0.3	0.5	0	C	4104	0
1983	11474	11695	2421		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	4372	0
1984	15400	24499	7460		0.6	0.8	0.6	0.8,	0.6	0.8	0.3	0.5	0	C	2935	0
1985	17779	45321	8296		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	3101	0
1986	13714	64351	11389		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	3464	0
1987	19641	56381	7087		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	5366	0
1988	13233	34200	9053		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	5523	0
1989	8736	20699	3592		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	4684	0
1990	1410	20055	5303		0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	C	3309	0
1991	9588	13336	1325		0.6	0.8	0.6	0.8	0.6	0.8	0.22	0.39	0	C	2323	0
1992	3893	12037	1144		0.6	0.8	0.6	0.8	0.6	0.8	0.13	0.25	0	C	2738	251
1993	3303	4535	802		0.6	0.8	0.6	0.8	0.6	0.8	0.10	0.19	0	C	2508	1793
1994	3202	4561	217		0.6	0.8	0.6	0.8	0.6	0.8	0.07	0.13	0	C	2549	3681
1995	1676	5308	865		0.6	0.8	0.6	0.8	0.6	0.8	0.04	0.07	0	C	2493	3302
1996	1728	8025	332		0.3557	0.4163	0.748	0.85	0.6	0.8	0.05	0.08	0	C	2565	3776
1997	9753				1	1	1	1	1	1	0.05	0.08	0	C	2365	2187
1998	0	0	0	2988	1	1	1	1	1	1	0.05	0.08	97408	205197	2131	3758
1999	0	0	0	2739	1	1	1	1	1	1	0.05	0.08	94894	199901	2076	4407
2000	0	0	0	5323		1	1	1	1	1	0.05	0.08	117063	246602	2561	7095
2001	0	0	0	4789	1	1	1	1	1	1	0.05	0.08	93660	197301	2049	4640
2002	0	0	0	5806	1	1	1	1	1	1	0.05	0.08	62321	142951	2071	5052
2003	0	0	0	6477	1	1	1	1	1	1	0.05	0.08	48256	122813	2112	4924
2004	0	0	0	8385	1	1	1	1	1	1	0.05	0.08	69808	120244	1808	5968
2005	0	0	0	10436	1	1	1	1	1	1	0.05	0.08	160038	281401	2007	7120
2006	0	0	0	10377	1	1	1	1	1	1	0.05	0.08	132205	294669	***************************************	5815
2007	0	\$	0	9208	1	1	1	1	1	1	0.05	0.08	131895	257360	1762	4641
2008	0	0	0	9834	1	1	1	1	1	1	0.05	0.08	142851	264694	1936	5917
2009	0	\$ -	0	7988	1	1	1	1	1	1	0.07	0.14	55307	149372	1355	3396
2010	0	0	0	9867	1	1	1	1	1	1	0.07	0.14	78560	165165	1477	4704
2011	0	\$ -	0	11138	1	1	1	1	1	1	0.07	0.14	137465	356791	1628	5340
2012	0	0	0	9977	1	1	1	1	1	1	0.07	0.14	105443	241754	1376	3302
2013	0	0	0	7185	1	1	1	1	1	1	0.07	0.14	83556	227719	1389	4167
2014	0	0	0	8958	1	1	1	1	1	1	0.07	0.14	175938	359832	1370	3797
2015	0	0	0	8923	1	1	1	1	1	1	0.07	0.14	174788	339699		3970

Appendix 4.vi. Input data for returns of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

		Salmon Fishin	g Area 3			Salmon Fis	shing Area	4		Salmon Fis	hing Area	5		Salmon Fi	shing Area	6		Salmon Fis	shing Area	a 7		Salmon Fis	hing Are	a 8	\neg
	Small	salmon	Large	salmon	Small	salmon	Large s	salmon	Small	salmon	Large s	almon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	П
	Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		
Year	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	
Bugs labe	SFA3Sm_L[]	SFA3Sm_H[]	SFA3Lg_L	SFA3Lg_F	SFA4Sm_	SFA4Sm_	SFA4Lg_L	SFA4Lg_F	SFA5Sm_	SFA5Sm_	SFA5Lg_L	SFA5Lg_F	SFA6Sm_	SFA6Sm_	SFA6Lg_L	SFA6Lg_	HSFA7Sm_	SFA7Sm_	SFA7Lg_	L SFA7Lg_	HSFA8Sm_	SFA8Sm_	SFA8Lg_	L SFA8L	<u>g_</u> F
1970	2613	5227	155	737	16163	32327	957	4559	7420	14840	439	2093	280	560) 17	79	9 67	133		4 1	9 62	123		4	17
1971		4947	146	698	12610	25220	746	3557	5600	11200	331	1579	183	367	7 11	5.	2 133	267		8 3		167		5	24
1972								3238	6317			1782			3 23			407		2 5				6	26
1973					22367			6308	7040	14080	417	1986	833						2					19	88
1974								4131	5457	10913		1258						887	5					20	39
1975								5587		13253		1869												11	82
1976								5365		12653		1524			$\overline{}$	198				2 2				32	64
1977								5290		30773		2908			126										51
1978								3461		19053		1128				_								9	17
1979								3067		8873		509			7 47	_								19	38
1980								2819		18013		809				_				5 3				18	36
1981								11996				3091		4033						5 10				34	68
1982								1604				391				4				2 2				7	14
1983								2740		15713		721												6	13
1984								5337				1457								5 5				4	43
1985								6352				1816			A					6 5				6	54
1986								4977				2170								, ,				12	74
1987								3078				939		1085						3 1				7	24
1988		13644						5367				2269		3333						2 7					43
1989		6114						2365		11740		885	_	2038						8 4				12	62
1990								3033				1293 1167								-				2	46 6
1991		9281						2788												_				0	0
1992 1993							1408 1605	13867	14189		516	5088	1681 2574	3363 4954										15	
								6242		32071	586	2280												7	58
1994								7343		21395		2433												1	28 48
1995 1996								8211		26762 38875	1201	2901 3806								4 6				16	48
1997								11497 8125			674	2183								1 3:				6	20
1998								18658		16673		4228								5 8				16	58
1999		20350						18651	10474	15245	964	3143								6 2				14	45
2000								7248		16734		2396								2 3				8	20
2001								7143	40007	13095		1791									9 20			2	4
2002								6556		5799		739		361							0 72			5	14
2003								12032				1062								5 1				3	8
2004								10464		13065		1769									0 41			2	9
2005		18899								15627	403	2214						-			0 26			2	11
2006								12872		13700		2581						-		-	0 172			20	
2007								13567				3208						-		5 4				2	52 6
2008		23285						15508	11459			2800								6 2				15	48
2009												4055									0 135			8	52
2010							6094				1766	5040						263	1	6 40				8	24
2011							5033	24165				5758								3 40					109
2012		33459						13600		23065		3045								3 2					77
2013								23377	17957	26461	1003	7859													55
2014							5170	13266	17550	26420	1374	3527													21
2015							_	20631	17154	28934	1413	4233													88

Appendix 4.vi. (Continued). Input data for returns of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

		Salmon	Fishi	ng Area 9		S	almon Fis	hing Area 10		Sal	mon Fish	ing Area 1	1	Sa	almon Fish	ning Area	12	S	almon Fis	hing Area	13		Salmon Fish	ing Area 14A	
	Small	salmor	ı	Large sali	mon	Small	salmon	Large salmo	n	Small sa	lmon	Large s	almon	Small s	salmon	Large	salmon	Small	salmon	Large	salmon	Smal	l salmon	Large	salmon
	Returns		R	eturns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns	
Year	Min	Max	M	in Ma	ìх	Min	Max	Min Max		Min N	lax	Min I	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
Bugs labe	SFA9Sm_	SFA99	Sm_S	FA9Lg_L SF	A9Lg_F	SFA10Sm	SFA10Sm	SFA10Lg_SFA1	0Lg_	SFA11Sm S	FA11Sm	SFA11Lg_	SFA11Lg_	SFA12Sm	SFA12Sm	SFA12Lg	SFA12Lg	SFA13Sm	SFA13Sm	SFA13Lg	SFA13Lg	SFA14ASm	SFA14ASm_H	SFA14ALg_L	SFA14ALg_F
1970	6310) 1:	2620	373	1780	2003	4007	119	565	16760	33520	992	4727	2497	4993	148	704								
1971	5400) 10	0800	320	1523	3093	6187	183	872	13533	27067	801	3817	1513	3027	90	427	26011	40151	2678	3 4750	12523	25047	308	2173
1972	3797	7 '	7593	225	1071	1890	3780	112	533	16350	32700	968	4611	3093	6187	183	872	23526	37589	3107	5169	8057	16113	198	1398
1973	7200) 1-	4400	426	2031	5950	11900	352	1678	16187	32373	958	4565	2153	4307	127	607	27287	40227	7 3303	5200	17607	35213	433	3055
1974	4980) !	9960	574	1149	4040	8080	466	932	14920	29840	1720	3441	2193	4387	253	506	19274	1 28824	1 2913	3 4257	7 10400	20800	902	1805
1975	6240) 1:	2480	880	1760	1423	2847	201	401	15003	30007	2116	4232	1700	3400	240	479	33671	54424	4497	7424	16060	32120	507	101:
1976	5410) 10	0820	651	1303	2433	4867	293	586	13880	27760	1671	3343	990	1980	119	238	29382	2 46902	2 3378	5488	3 24603	49207	1437	287
1977	3600) '	7200	340	680	3657	7313	346	691	13653	27307	1290	2581	1860	3720	176	352	17610	25240	2877	7 3598	3 19023	38047	666	133
1978	3 4343	3	8687	257	514	5317	10633	315	629	13320	26640	788	1576	1220	2440	72	144	17807	7 27681	4716	5289	10803	21607	266	533
1979	5680) 1	1360	326	651	2830	5660	162	324	11433	22867	655	1311	2443	4887	140	280	20372	31829	1183	1862	2192	7 43853	233	46
1980	7930) 1:	5860	356	712	5080	10160		456	16897	33793	759	1518	2733	5467	123	246	26538	38871	5236	5 5913	1247	7 24953	694	
1981			2413	825	1650				1167	23540	47080	3129	6258			470		31359						1090	
1982			2167	147	293				202	24460	48920	590	1180					31628							
1983	7677	7 1:	5353	352	705	3800	7600	174	349	15897	31793	730	1460	2223	4447	102	204	20828	31701	1 4465	5100	1266	25333	1704	340
1984	7989) 1'	7023	125	1221	5141	10955	81	785	24767	52774	389	3784	6782	14451	106	1036	26184	37852	2 2296	3710	16962	36143	266	259
1985	6375	5 1:	3198	96	912	4831	10000	73	691	21213	43914	320	3034	3996	8273	.60	572	16028	3 25505	1375	5 2508	3 13209	27345	199	1890
1986	8411	1 1	7555	208	1231	5619	11727	139	822	20300	42368	501	2970	3433	7166	85	502	22881	36916	5 2079	3649	1841	38426	455	2694
1987	3416	5 (6865	88	489	1690	3397	44	242	15087	30317	391	2162	3274	6580	85	469	19629	32325	1546	3022	18203	36580	471	2608
1988	5179) 10	0668	124	748	4308	8873	103	622	18985	39106	456	2741	5330	10979	128	770	26162	2 43480	1950	3917	7 23580	48570	566	3405
1989	5352	2 10	0895	160	821	3655	7440	109	561	12047	24524	360	1849	2279	4640	68	350	10154	16156	5 849	1565	13036	26537	390	200
1990	7332	2 1:	2834	256	1000	3281	5743	115	447	17470	30578	610	2382	3363	5887	117	459	21518	31183	1778	3084	19843	34732	693	2700
1991	2404	1 :	3949	82	319	988	1622	2 34	131	7956	13068	272	1056	2765	4542	95	367	16225	20945	1709	2433	15307	25141	523	203
1992	5044	1 10	0088	184	1809	1791	3582	2 65	642	16615	33231	605	5958	4671	9342	170	1675	25990	44119	3087	7 8928	3492	7 69854	1271	12525
1993	3 11402	2 2	1948	401	1560	5578	10736	196	763	24574	47301	865	3363	5936	11426	209	812	27523	46889	2618	3 4746	31116	59893	1095	
1994	1 3007	7 (6607	177	751	2544	5588	150	635	7649	16803	450	1910	2761	6066	162	690	22103	37166	3476	5879	1332	29263	783	332
1995	5 5321	1 1:	2821	318	1390	4371	10532	261	1142	10757	25916	642	2809	2294	5527	137	599	27022	2 49781	1843	5096	5 20840	50209	1244	5443
1996	6015	5 1:	3450	443	1317	8245	18438	608	1805	18938	42350	1396	4146	5025	11238	370	1100	36576	67672	3479	7132	2 3276	73263	2415	7172
1997	3636	5 (6627	379	1227	5071	9242	528	1712	16648	30339	1735	5619	4556	8303	475	1538	31402	2 46494	4240	8521	2524	45998	2630	8519
1998	3 4694	1	6597	468	1673	7821	10992	780	2788	8467	11900	844	3018	2360	3318	235	841	21816	5 27955	3194	1 7080	23995	33724	2392	855
1999	4015	5	5844	369	1205	5113	7443	3 470	1535	9643	14036	887	2894	1139	1658	105	342	32407	40858	3878	3 7739	26960	39241	2480	809
2000	7850) 1	0582	596	1515	7639	10297	580	1475	17260	23266	1310	3332	2634	3551	200	509	54330	67784	5519	10048	36819	49632	2795	710
2001	2043	3	2674	158	366	2924	3826	5 226	523	9396	12296	725	1682	2201	2880	170	394	37393	3 45761	3749	6510	20775	27188	1604	
2002	1917	7 :	2873	125	366	3713	5565		709	9011	13505	587	1722	2321	3478	151	443	34070	46011	3452	2 6469	26558	39801	1729	5075
2003	2229) :	2699	113	359	3771	4565	190	608	14208	17201	718	2291	5917	7163	299	954	50367	57997	7 4421	8434	40802	49395	2061	6579
2004	1926	5	3001	95	406	3697	5760	183	780	13762	21443	681	2903	3131	4879	155	661	49924	66549	4308	9118	30057	46833	1488	634
2005	1948	3	5734	148	813	2779	8180	211	1159	6260	18425	475	2611	2686	7905	204	1120	40658	88340	4595	12966	5 17340	51040	1316	7232
2006	4355	5 (6960	506	1311	5344	8542		1609	11033	17634	1283	3322	3460	5530	402	1042	53311	74546	5 8499	15058	3 2808	44883	3266	8450
2007			4817	270	873				1285	5650	11449	641	2076												
2008	3944	1 (6606	301	963	4786	8016	366	1169	11136	18654	851	2721	2610	4373	200	638	51933	75122	2 3901	9668	3 25802	2 43220	1972	630
2009	3445	5 (6443	203	1316	5137	9608	303	1963	7536	14097	445	2880	1746	3266	103	667	36368	55458	3722	2 10806	5 21146	39555	1249	808
2010	6597	7 8	3227	504	1440	8168	10187	625	1783	8024	10008	614	1751	2999	3740	229	654	57930	67116	5 5798	11067	31675	39504	2422	691
2011			3983	438	2105				3600	6897	11755	574	2754		4243				68766	3356	16112	2 24110	41092	2005	
2012			9539	474	1259				1579	6727	9554	475	1261	2624	3726										
2013	4760	7	015	266	2083	8060	11876	450	3527	7372	10863	412	3226	2043	3011	114	894	33752	49737	3478	13815	13924	20518	778	609
2014			749	403	1034				1193	4408	6637	345	886		2640										
2015	4795	5 8	8089	395	1183				1205	6057	10217	499	1495		3655								54756	2674	

Appendix 4.vii. Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

		Salmon Fishin	g Area 3			Salmon F	ishing Area	1		Salmon Fis	shing Area	5		Salmon Fi	shing Area	6		Salmon F	ishing Area	7		Salmon Fi	ishing Area	8
	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon
	Spawners		Spawners		Spawners		Spawners		Spawners		Spawners	_	Spawners		Spawners		Spawners		Spawners		Spawners	:	Spawners	3
Year	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
	SFA3SSm L[]	SFA3SSm H	1 SFA3SLq	SFA3SLq	SFA4SSn	SFA4SS	m SFA4SLq	SFA4SLq	SFA5SSm	SFA5SSn	SFA5SLo	SFA5SLq	SFA6SSn	n SFA6SSi	n SFA6SLo	SFA6SLc	SFA7SSr	n SFA7SSi	m SFA7SLq	SFA7SLc	SFA8SSr	n SFA8SSr	m SFA8SLc	SFA8SLo
1970	1829	4443	154	736	11314	27478	910	4512	5194	12614	404	2058	196	476	14	76	47	113	3	18	43	105	0	13
1971	1731	4205	135	687	8827	21437	688	3499	3920	9520	293	1541	128	312	10	51	93	227	8	38	58	142	0	15
1972	1162	2822	98	468	8036	19516	655	3214	4422	10738	354	1762	278	674	23	112	142	346	12	57	65	159	6	26
1973	2772	6732	232	1115	15657	38023	1275	6259	4928	11968	405	1974	583	1417	49	235	306	742	26	123	219	533	15	84
1974	1958	4754	318	641	12537	30447	1983	4049	3820	9276	608	1237	707	1717	115	232	310	754	49	100	119	289	20	39
1975	2583	6273	520	1041	13867	33677	2628	5421	4639	11265	912	1846	219	533	43	87	93	227	19	38	203	493	41	82
1976	2210	5366	379	759	15594	37870	2495	5177	4429	10755	697	1459	576	1400	97	196	70	170	12	24	187	453	32	64
1977	3570	8670	478	960	19591	47577	1559	4204	10771	26157	1410	2864	936	2272	107	234	182	442	24	48	189	459	26	51
1978	1769	4295	149	298	20473	49719	1229	2959	6669	16195	536	1100	691	1677	51	110	231	561	19	38	103	249	9	17
1979	4760	11560	390	779	18727	45481	1206	2740	3106	7542	234	489	569	1383	45	91	292	708	24	48	233	567	19	38
1980	4067	9877	224	485	21966	53346	903	2312	6305	15311	376	780	747	1813	34	82	238	578	14	30	280	680	18	36
1981	5502	13362	1042	2087	31584	76704	5637	11635	8139	19765	1511	3056	1412	3428	239	507	287	697	53	107	180	436	34	68
1982	6146	14926	124	336	23270	56514	544	1346	5677	13787	143	338	672	1632	6	29	362	878	2	15	198	482	0	5
1983	3773	9163	245	493	20893	50739	1073	2443	5500	13356	191	551	691	1677	35	81	324	788	0	9	96	232	1	8
1984	2531	6525	55	540	25033	64536	533	5322	6835	17620	149	1456	789	2034	12	163	243	626	1	48	200	515	4	43
1985	3462	8569	72	683	32218	79741	671	6352	9208	22791	192	1816	1134	2806	24	224	296	733	6	58	272	674	6	54
1986	2054	5126	70	413	24722	61700	840	4977	10782	26910	366	2170	1184	2955	40	238	271	677	9	55	367	916	12	74
1987	1655	3895	57	318	16032	37722	556	3078	4892	11511	170	939	403	948	14	77	82	194	3	16	126	297	4	24
1988	4868	11888	159	956	27317	66712	892	5367	11549	28204	377	2269	1189	2904	39	234	355	867	12	70	219	535	7	43
1989	2266	5376	90	461	11623	27581	461	2365	4350	10323	172	885	755	1792	30	154	203	481	8	41	304	721	12	62
1990	5032	10098	236	920	16583	33273	776	3033	7071	14188	331	1293	978	1963	46	179	144	288	7	26	252	505	12	46
1991	4334	7965	193	750	16113	29607	718	2788	6745	12395	301	1167	613	1126	27	106	119	218	5	21	36	67	2	6
1992	9844	21262	415	4094	33228	71898	1407	13866	12175	26363	516	5088	1450	3132	61	603	252	545	11	105	0	0	0	0
1993	10054	20961	400	1599	39162	81344	1590	6226	14370	29779	576	2270	2243	4623	90	351	404	831	16	63	369	760	15	58
1994	9146	24802	749	3247	20576	55760	1644	7259	6855	18510	560	2420	381	1026	30	133	46	122	4	16	79	212	6	27
1995	7409	21791	580	2636	22872	67179	1801	8135	8122	23776	642	2880	287	831	23	100	173	501	14	60	135	397	11	48
1996	15729	39860	1412	4247	42346	107268	3757	11383	14095	35586	1263	3787	522	1317	46	139	124	311	11	33	180	457	16	48
1997	9422	19095	1209	3954	19309	39107	2467	8083	5228	10547	668	2177	190	384	24	79	49	99	6	20	48	98	6	20
1998	16390	24345	1933	6969	43559	64785	5160	18599	9943	14753	1155	4201	455	673	53	191	212	313	25	88	135	201	16	57
1999	11804	18173	1279	4189	52390	80698	5650	18583	8832	13603	947	3126	343	528	37	121	58	90	6	21	119	188	14	45
2000	17003	23723	1449	3711	32879	45946	2803	7201	10897	15217	923	2377	993	1386	84	217	140	195	12	31	88	125	8	20
2001	9861	13489	892	2089	33365	45682	3023	7086	8344	11433	767	1786	250	342	23	53	42	59	4	9	17	23	2	4
2002	8620	13856	671	1994	28099	45208	2175	6498	3194	5124	250	737	199	319	15	45	0	0	0	0	55	91	5	14
2003	19386	23938	1085	3478	67026	82739	3738	12001	5926	7312	331	1060	412	508	23	74	94	116	5	17	47	58	3	8
2004	6942	11402	390	1680	43104	70785	2430	10438	7307	11987	412	1766	158	259	9	38	0	0	0	0	35	58	2	9
2005	5056	17534	473	2664	28896	100323	2695	15235	4200	14518	394	2205	92	314	8	47	0	0	0	0	18	69	2	11
2006	9402	15839	1228	3216	37156	62732	4925	12825	7495	12623	969	2554	61	102	8	20	0	0	0	0	141	244	20	52
2007	9147	19842	1171	3818	32243	70143	4122	13501	7641	16603	978	3196	68	148	8	28	112	245	12	45	15	33	2	6
2008	11799	21183	1045	3379	53591	96443	4745	15402	9669	17405	867	2791	274	497	22	78	69	125	4	18	159	292	15	48
2009	11205	22795	779	5080	49881	101728	3491	22732	8828	18065	622	4049	412	834	28	185	0	0	0	0	111	228	7	51
2010	18364	23569	1595	4581	69075	88772	6006	17304	20114	25822	1754	5028	874	1120	76	217	183	235	16	46	93	120	8	24
2011	13193	24264	1291	6261	50806	93428	4789	23920	12075	22230	1176	5734	716	1314	70	339	83	153	8	39	220	412	22	108
2012	21149	31048	1639	4394	64959	95436	5046	13528	14554	21377	1140	3039	738	1086	57	154	100	147	8	21	361	533	25	73
2013	7822	12219	495	4039	44838	70136	2889	23284	15027	23531	976	7833	721	1128	44	373	241	379	16	127	102	162	7	55
2014	11442	18452	1074	2775	54396	87757	5068	13164	14519	23389	1345	3497	693	1115	62	165	255	416	24	63	83	134	8	21
2015	18245	32560	1690	5117	73398	130811	6786	20531	15040	26820	1392	4213	1019	1814	94	284	238	429	23	68	301	546	29	88

Appendix 4.vii. (Continued). Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

		Salmon Fis	shing Area 9	9	,	Salmon Fis	hing Area 1	0		Salmon Fisl	ning Area 1	1		Salmon Fis	shing Area	12	5	Salmon Fis	hing Area	13		Salmon Fisl	hing Area 14A	•
	Small s	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small s	salmon	Large	salmon	Sma	II salmon	Large	salmon
	Spawners		Spawners		Spawners		Spawners		Spawners		Spawners		Spawners		Spawners		Spawners		Spawners	3	Spawners		Spawners	
Year	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
[]	SFA9SSm	SFA9SSn	SFA9SLg	SFA9SLg	SFA10SS	SFA10SS	SFA10SL	SFA10SL	SFA11SS	SFA11SS	SFA11SL	SFA11SL	SFA12SS	SFA12SS	Sr SFA12SL	SFA12SL	SFA13SSi	SFA13SS	SFA13SL	SFA13SL	gSFA14ASSi	m SFA14ASSm	SFA14ASLg	SFA14ASL
1970	4417	10727	361	1768	1402	3406	112	558	11732	28492	918	4653	1748	4244	69	625	16203	28543	1608	3417	10372	25188	134	2340
1971	3780	9180	301	1504	2165	5259	166	855	9473	23007	736	3752	1059	2573	74	411	16489	30629	1633	3705	8766	21290	0	1850
1972	2658	6454	217	1063	1323	3213	108	529	11445	27795	882	4525	2165	5259	163	852	15125	29188	2004	4066	5640	13696	83	1283
1973	5040	12240	406	2011	4165	10115	310	1636	11331	27517	923	4530	1507	3661	102	582	17019	29959	1911	3808	12325	29931	91	2713
1974	3486	8466	565	1140	2828	6868	452	918	10444	25364	1682	3403	1535	3729	240	493	12085	21635	1997	3341	7280	17680	789	1692
1975	4368	10608	874	1754	996	2420	192	392	10502	25506	2076	4192	1190	2890	220	459	21668	42421	3611	6538	11242	27302	417	925
1976	3787	9197	639	1291	1703	4137	283	576	9716	23596	1629	3301	693	1683	114	233	18999	36519	2752	4862	17222	41826	1337	2774
1977	2520	6120	331	671	2560	6216	341	686	9557	23211	1272	2563	1302	3162	128	304	10898	18528	1828	2549	13316	32340	194	859
1978	3040	7384	240	497	3722	9038	273	587	9324	22644	770	1558	854	2074	52	124	12518	22392	3861	4434	7562	18366	194	460
1979	3976	9656	311	636	1981	4811	154	316	8003	19437	648	1304	1710	4154	130	270	14363	25820	1070	1749	15349	37275	174	408
1980	5551	13481	295	651	3556	8636	201	429	11828	28724	715	1474	1913	4647	94	217	18625	30958	4243	4920	8734	21210	514	1208
1981	4345	10551	773	1598	3073	7463	555	1138	16478	40018	3088	6217	2473	6007	453	922	22059	36689	4485	6789	13725	33331	953	2043
1982	4258	10342	114	260	2931	7117	91	192	17122	41582	537	1127	3628	8812	140	235	22062	37132	2847	3236	11114	26990	2987	6082
1983	5374	13050	281	634	2660	6460	95	270	11128	27024	703	1433	1556	3780	94	196	14491	25364	3855	4490	8867	21533	1635	3338
1984	5725	14759	120	1216	3684	9498	79	783	17748	45755	374	3769	4860	12529	38	968	18413	30081	1987	3401	12155	31336	179	2504
1985	4625	11448	96	912	3505	8674	73	691	15390	38091	320	3034	2899	7176	57	569	10726	20203	1349	2482	9583	23719	197	1887
1986	6113	15257	208	1231	4084	10192	139	822	14754	36822	501	2970	2495	6228	81	499	15535	29570	2013	3583	13381	33396	445	2683
1987	2549	5998	88	489	1261	2968	44	242	11258	26488	391	2162	2443	5749	87	466	13611	26307	1512	2988	13583	31960	467	2604
1988	3806	9295	124	748	3166	7731	103	622	13952	34073	456	2741	2017	9566	126	767	17945	35263	1909	3877	17329	42319	549	3388
1989	4037	9580	160	821	2757	6542	103	561	9087	21564	360	1849	1719	4080	67	349	6980	12982	836	1552	9833	23334	385	1996
1990	5466	10968	256	1000	2446	4908	115	447	13024	26132	610	2382	2507	5031	114	456	14866	24531	1744	3051	14793	29682	679	2692
1990	1844	3389	82	319	758	1392	34	131	6103	11215	272	1056	2121	3898	93	365	11037	15757	1689	2413	11742	21576	512	2092
1992	4334	9378	183	1809	1496	3287	65	642	14239	30854	605	_		8657	162	1667	20506	38635	2992	8833	30096	65023	1234	12488
1992	9956	20502	400	1559	4809	9967	194	761	21423	44150	861	5958 3359	3985 5176	10666	207	810	22341	41708	2544	4673	27010	55787	1058	4221
		5723		746	1804		194	630		14449	430	1891	1949	5253	154	681	15381		3207		9385			3286
1994	2124		172			4848			5295									30444		5611		25327	742	
1995	3887	11386	304	1376	3218	9378	253	1133	7770	22930	625	2792	1689	4922	130	592	20570	43329	1607	4860	15218	44587	1187	5385
1996	4868	12304	431	1304	6687	16880	592	1789	15226	38638	1362	4113	4082	10295	358	1088	29056	60152	3199	6852	26584	67085	2357	7115
1997	2927	5918	372	1221	4086	8257	519	1702	13304	26995	1718	5602	3655	7401 2925	464	1527	25508	40599	3985 3031	8266	20359	41117	2578 2347	8467
1998	3937	5840	458	1663	6606	9777	771	2779	7024	10457	8.50	3009	1968		225	831	18279	24417		6918	19992	29721		8507
1999	3401	5230	359	1195	4313	6642	455	1520	8086	12478	881	2889	958	1477	102	339	28647	37098	3760	7621	22659	34941	2402	8013
2000	6913	9645	581	1501	6664	9322	534	1429	14895	20901	1288	3310	2291	3208	195	504	48055	61508	5250	9779	32314	45127	2731	7044
2001	1709	2339	151	359	2436	3338	215	513	7804	10704	714	1671	1818	2497	162	386	31037	39405	3536	6297	17331	23744	1559	3674
2002	1562	2518	118	360	3049	4901	231	699	7347	11840	581	1716	1896	3053	147	439	28083	40025	3313	6330	21764	35007	1668	5013
2003	1985	2454	109	355	3368	4162	185	603	12701	15693	703	2276	5282	6528	288	943	45027	52657	4206	8218	36597	45189	1988	6506
2004	1674	2749	91	402	3210	5273	177	774	11863	19544	660	2882	2704	4452	149	655	43889	60513	4074	8883	26116	42892	1429	6282
2005	1478	5264	130	794	2171	7572	194	1142	4827	16992	456	2591	2062	7282	191	1107	33349	81031	4320	12691	13676	47376	1246	7163
2006	3791	6397	498	1302	4627	7824	602	1590	9554	16155	1271	3310	2986	5056	392	1032	46296	67532	8247	14807	24532	41334	3210	8400
2007	2063	4502	263	867	3047	6636	387	1275	4907	10706	636	2071	2442	5323	314	1027	29402	54734	4511	10780	17446	37934	2222	7293
2008	3285	5948	293	955	3971	7202	351	1154	9314	16832	841	2711	2178	3940	193	631	43277	66465	3580	9346	21887	39305	1915	6246
2009	2835	5834	198	1311	4193	8665	298	1957	6203	12763	442	2877	1450	2970	100	664	31106	50196	3526	10610	17820	36229	1200	8032
2010	5703	7334	496	1432	7062	9081	616	1774	6859	8842	604	1742	2606	3347	226	651	49703	58889	5478	10747	27468	35298	2358	6848
2011	4364	8077	433	2099	7477	13826	716	3566	5696	10554	564	2744	2074	3827	203	990	33849	62267	3160	15915	20249	37231	1953	9575
2012	5898	8720	471	1256	7488	11027	581	1566	5993	8819	468	1255	2348	3450	184	490	44778	65820	3395	9251	31467	46268	2451	6571
2013	3973	6228	254	2071	6681	10498	424	3502	6130	9621	398	3213	1701	2668	104	885	28314	44299	3301	13638	11746	18341	734	6050
2014	4174	6776	400	1031	4862	7861	431	1158	3602	5830	340	881	1445	2331	134	349	18039	29192	2707	5849	14639	23622	1361	3541
2015	4169	7462	385	1173	4265	7618	389	1192	5250	9410	489	1485	1891	3379	167	524	43908	78364	5419	13980	28414	50707	2619	7956

Appendix 4.viii. Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

		Salmon Fisl	ning Area 3			Salmon Fisl	ning Area 4			Salmon Fis	hing Area 5			Salmon Fish	ning Area 6		Salmon Fis	shing Area 7		Salm	on Fish	ing Area 8	
	2	SW		2SW	28	W	2SV	٧	2SV	V	2SW		281	W	2SW	25	SW	25	SW	2SW		28	W
	Returns		Spawners		Returns		Spawners		Returns		Spawners		Returns		Spawners	Returns		Spawners		Returns	5	Spawners	
Year	Mir	Ma:	x Mi	n Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min Ma	x Mir	n Max	Min	Max	Min	Max	Min	Max
Bugs labels	SFA3R2 L[]	SFA3R2 H[]	SFA3S2 L	SFA3S2 H[]	SFA4R2 L[] S	SFA4R2 H[]	SFA4S2 L[] S	FA4S2 H[]	SFA5R2 L[]S	FA5R2 H[]	SFA5S2 L[] SF	A5S2 H[]	SFA6R2 L S	SFA6R2 H[SFA6S2_L SFA6S2_H	SFA7R2 I	SFA7R2 H	SFA7S2 L	SFA7S2 H	SFA8R2 L SFA8	8R2 H S	SFA8S2 L	SFA8S2 H
1970	1:					912	91	902	44	419		412		16		15 (0	3	0	
1971	1:	14	0 1	4 137	7 75	711	69	700	33	316	29	308	1	10	1 1	10 1	1 8	1	8	0	5	0	3
1972	10	9	4 1	0 94	4 68	648	66	643	37	356	35	352	2	22	2 2	22 1	1 11	1	11	1	5	1	5
1973	2:	3 22	3 2	3 223	3 132	1262	127	1252	42	397	40	395	5	47	5 4	17 3	3 25	3	25	2	18	1	17
1974	3:	12	9 3	2 128	3 207	826	198	810	63	252	61	247	12	47	12 4	16 5	5 20	5	20	2	8	2	8
1975	50	20	8 5	2 208	3 279	1117	263	1084	93	374	91	369	4	18	4 1	17 2	2 8	2	8	4	16	4	16
1976	31	15	2 3	8 152	2 268	1073	249	1035	76	305	70	292	10	40	10	189	1 5	1	5	3	13	3	13
1977	4	19	3 4	8 192	2 264	1058	156	841	145	582	141	573	13	51	11 4	17 2	2 10	2	10	3	10	3	10
1978	13	6	0 1:	5 60	173	692	123	592	56	226		220		23	5 2	2 2		_		1	3	1	3
1979	39					613	121	548		102		98		19		8 2			10		8	2	
1980	20					564	90	462		162		156	5	19		6 2	- 0		6	2	7	2	
1981	10-					2399	564	2327	155	618		611	27	107	24 10		,			3	14	3	
1982	2					321	54	269	20	78		68	2	9	-	6 1	1 5			1	3	0	_
1983	2:					548	107	489	36	144		110		18		6 2	2 9		_	1	3	0	_
1984		5 10		6 108		1067	53	1064	15	291	15	291	2	34		13 1	1 10				9	0	
1985		13		7 137		1270	67	1270	19	363		363	2	45		15 1	1 12		12		11	1	11
1986		8		7 83		995	84	995	37	434		434	4	48		18 1	1 11		11		15	1	
1987		5 6		6 64		616	56	616		188		188	1	15		5 (, .			0	5	0	-
1988	10			6 191		1073	89	1073		454		454	4	47		17 1	1 14		14	1	9	1	9
1989				9 92		473	46	473		177		177	3	31		81 1	1 8		8	1	12	1	12
1990	2-					607	78	607	33	259		259	5	36		36 1	1 5	_	5	1	9	- 1	9
1991	19					558	72	558		233	30	233	3	21		21 1	1 4		4	0	1	0	-
1992	4:					2773	141	2773		1018	52	1018		121	6 12		1 21		21	0	0	0	
1993	4:					1248	159	1245		456	58	454		70		-				0	12	- 1	12
1994	3					1028	99	1016		341	34	339		19) 2		2	0	7	0	4
1995 1996	8					1150 1610	108 225	1139 1594	77	400	76	403 530		14 20		14 1	1 9		5	1	7	1	7
1997	74					1138	148	1132		393	40	305		11		1 (-	0	3	0	3
1998	111					2612	310	2604	71	592	- 10	588		27		27 1	1 12		12	1	8	1	3
1999	7					2611	339	2602		392 440	57	438		17		7 (1	6	1	6
2000						1015	168	1002	57	225	55	333		30		80 1	1 4		J	0	3	0	2
2001	3					664	130	659	33	167		166		5		5 (0	1	0	0	0	
2001	2					610	94	604		69		69		4		4 (-		0	1	0	
2003	4					1119	161	1116	14	99		99		7	1	7 (0	1	0	-
2004						973	104	971	18	165		164		4	0	4 () (0	1	0	-
2005	2					1426	116	1417		206		205		4	0	4 () (0	0	0	1	0	1
2006	5-					1197	212	1193	43	240		237		2	0	2 () (0	0	1	5	1	5
2007	5					1262	177	1256		298		297		3	0	3 1	1 4		4	0	1	0	1
2008						1442	204	1432		260		260		7	1	7 () 2	0	2	1	4	1	4
2009	3-					2117	150	2114	27	377		377		17	1 1	7 (0	0	5	0	5
2010						1618	258	1609		469		468		20		20 1	1 4	1	4	0	2	0	2
2011	56					2247	206	2225	52	535		533	-	32	3 3) 4	0	4	1	10	1	10
2012						1265	217	1258	49	283	49	283	3	14		4 () 2	. 0	2	1	7	1	7
2013						2174	124	2165	43	731	42	728		35		5 1	1 12		12	0	5	0	
2014						1234	218	1224	59	328		325		16		5 1			6	0	2	0	
2015						1919	292	1909	61	394	60	392		27		16 1			6	1	8	1	
_010				.,		.510		.500		50 1			-								-		

Appendix 4.viii. (Continued). Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

		Salr	non Fish	ing Area 9			Salmon Fi	shing Area 10	•———	·	Salmon Fish	ning Area 11			Salmon Fis	hing Area 12		<u> </u>	Salmon Fish	ning Area 13			Salmon Fish	ing Area 14A	
		2SW		25\	W		2SW	2	SW	2\$	W	2SW		281	N	2SV	٧	2SW	1	25	SW	2	SW	25	SW
	Returns	s	5	Spawners		Returns		Spawners		Returns		Spawners		Returns		Spawners		Returns		Spawners		Returns		Spawners	
Year		Min	Max	Min	Max	Mi	n Ma	x Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
Bugs labels	SFA9R	R2_L[] SFA9F	R2_H[] S	SFA9S2_L[] S	SFA9S2_H[]	SFA10R2_	L SFA10R2_F	[SFA10S2_L	SFA10S2_H[SFA11R2_L	SFA11R2_H[SFA11S2_L SFA	A11S2_H[SFA12R2_S	FA12R2_F	SFA12S2_S	FA12S2_F	SFA13R2_L[] S	FA13R2_H[]	SFA13S2_L[]	SFA13S2_H[]	SFA14AR2_L	SFA14AR2_H	SFA14AS2_L	SFA14AS2_H
1970		37	356	36	354	1	2 11	3 11	112	99	945	92	931	15	141	7	125	1300	3036	643	2050	36	514	13	468
1971		32	305	30	301	1	8 17	4 17	171	80	763	74	750	9	85	7	82	1071	2850	653	2223	31	435	0	370
1972		22	214	22	213	1	1 10	7 11	106	97	922	88	905	18	174	16	170	1243	3101	802	2439	20	280	8	
1973		43	406	41	402	3	5 33	6 31	327	96	913	92	906	13	121	10	116	1321	3120	764	2285	43	611	9	543
1974		57	230	57	228	4	7 18	6 45	184	172	688	168	681	25	101	24	99	1165	2554	799	2005	90	361	79	338
1975		88	352	87	351	2	0 8	0 19	78	212	846	208	838	24	96	22	92	1799	4454	1445	3923	51	203	42	185
1976		65	261	64	258	2	9 11	7 28	115	167	669	163	660	12	48	li	47	1351	3293	1101	2917	144	575	134	555
1977		34	136	33	134	3	5 13	8 34	137	129	516	127	513	18	70	13	61	1151	2159	731	1530	67	266	19	172
1978		26	103	24	99	3	1 12	6 27	117	79	315	77	312	7	29	5	25	1886	3173	1544	2660	27	106	19	92
1979		33	130	31	127	1	6 6	5 15	63	66	262	65	261	14	56	13	54	473	1117	428	1049	23	93	17	82
1980		36	142	30	130	2	3 9	1 20	86	76	304	71	295	12	49	9	43	2094	3548	1697	2952	69	278	51	242
1981		83	330	77	320	5	8 23	3 55	228	313	1252	309	1243	47	188	45	184	2059	4471	1794	4073	109	436	95	409
1982		15	59	11	52	1	0 4	0 9	38	59	236	54	225	13	50	N	47	1377	2298	1139	1941	309	1238	299	1216
1983		35	141	28	127	1	7 7	0 10	54	73	292	70	287		41	9	39		3060	1542	2694	170	681	163	668
1984		13	244	12	243		8 15	7 8	157	39	757	37	754	11	207	4	194	918	2226	795	2041	27	518	18	501
1985		10	182	10	182		7 13	8 7	138	32	607	32	607	6	114		114	550	1505	540	1489	20	378	20	377
1986		21	246	21	246	1	4 16	4 14	164	50	594	50	594	8	100	8	100	832	2190	805	2150	45	539	44	537
1987	·	9	98	9	98		4 4	8 4	48	39	432	39	432	8	94	8	93	618	1813	605	1793	47	522	47	521
1988		12	150	12	150	1	0 12	4 10	124	46	548	46	548	13	154	13	153	780	2350	764	2326	57	681	55	678
1989		16	164	16	164	1	1 11				370		370	7	70		70		939	334	931			39	
1990		26	200	26	200	1	1 8	9 11	89	61	476	61	476	12	92		91	711	1851	698	1830	69	541	68	538
1991		8	64	8	64		3 2	6 3	20		211	27	211	9	73		73		1460	676	1448				
1992		18	362	18	362		7 12	8 6	128	60	1192	60	1192	17	335	16	333	1235	5357	1197	5300	127	2505	123	2498
1993		40	312	40	312	2	0 15	3 19	152	86	673	86	672	21	162	21	162	1047	2848	1018	2804	110	852	106	
1994		11	105	10	104		9 8		- 00		267	26	265	10	97		95		3528	1283	3366				
1995		19	195	18	193		6 16				393		391	8	84	-	83		3058	643					
1996		27	184	26	183		6 25				580		576		154		152		4279	1280					
1997		23	172	22	171		2 24				787	103	784		215		214		5113						
1998		28	234	27	233		7 39				422	50	421	14	118		116		4248	1212					
1999		22	169	22	167		8 21				405	53	404		48		48		4643	1504					
2000		36	212	35	210		5 20				466	77	463		71		71		6029	2100					
2001		7	34	7	33		0 4		10		156	31	155		37		36		2324	658					
2002		5	34	5	33		0 6				160	25	160		41		41		2309	616					
2003		5	33	5	33		8 5		50	_	213		212		89		88		3011	782					
2004		4	38	4	37			3 8			270		268		61		61		3255	758					
2005		6	76	6	74		9 10		-	20	243		241		104		103		4629	804					
2006		22	122	21	121		7 15			55	309		308		97		96		5376						
2007		12	81	11	81		7 11			28	193		193		96		95		3912	839					
2008		13	90	13	89		6 10		_	37	253		252		59		59		3451	666					
2009		9	122	9	122		3 18			19	268		268		62		62		3858	656					
2010		22	134	21	133		7 16	_	165		163		162		61		61		3951	1019					
2011		19	196	19	195	3:			332		256		255		92		92		1498	136					
2012		20	117	20	117		6 14				117		117		46	-	46		873	146					
2013		11	194	11	193		9 32		326		300		299		83		82		3687	814					
2014		17	96	17	96	2			108		82		82		33		32		2482	741					
2015	4	17	110	17	109	1	7 11	2 17	111	21	139	21	138	8	50	- /	49	1219	4526	1169	4446	115	745	113	740

Appendix 4.ix. Input data for small salmon returns to Québec by category of data used in the run-reconstruction.

	Small returns								Small return	ns							
	Minimum			-					Maximum			-					
Year	C1	C2	C3	C4	C5			Other rivers			C3	C4	C5		FN Harvest Other river		Commercial
_		CSMC2_L[]	QCSmC3_L[]Q	(CSMC4_L[] C						-		QCSMC4_H[]Q	CSMC5_H[]C	.JH_dOMCOX .0	QCSmFn_H[] QCSmO_H		QCCmSm[]
1970	0	0	0	0	0	0	0	-			0	0	0	0		•	0 0
1971	-	0	0	0	0	0	0		-		0	0	0	0	-	•	0 0
1972		0	0	0	0	0	0	-	-	-	0	0	0	0		•	0 0
1973 1974		0	0	0	0	0	0	-			0	0	0	0	-	•	0 0
1974		0	0	0	0	0	0	-		Ü	0	0	0	0	-	-	0 0
1975	-	0	0	0	0	0	0	-			0	0	0	0	-	•	0 0
1976		0	0	0	0	0	0	-	-	-	0	0	0	0	-	•	0 0
1978		0	0	0	0	0	0	-	_	-	0	0	0	0	-	-	0 0
1976		0	0	0	0	0	0				0	0	0	0		•	0 0
1980		0	0	0	0	0	0				0	0	0	0		•	0 0
1980	0	0	0	0	0	0	0		-		0	0	0	0		U .	0 0
1982	-	0	0	0	0	0	0				0	0	0	0		0	0 0
1983		0		0	0	0	0				0	0	0	0	•	0	0 0
1984		5434	2955	460	1670	5160	267				6053	792	2784	8599	-	2 35	
1985		2271	1767	210	5449	4384	267				3586	352	9224	7307		7 27	
1986		5193	2396	63	6719	5133	267				4895	107	11198	8555			
1987		4775	3852	327	8396	5501	267				7875	546	13993	9168			
1988		5968	4404	468	8440	6423	267				8962	780	14067	10705			
1989		4743	2924	301	6744	5622	267		_		5940	503	11240	9369			
1990		7332	4377	694	7096	2976	377		13278		8917	1158	11826	4960			
1991		5851	3776	349	5009	2001	256				7679	584	8348	3336		5 26	
1992		6928	4567	428	5131	3462	243				9297	715	8552	5770			
1993		6325	3973	1029	4315	1447	525				8075	1717	7192	2412			7 3627
1994		5928	3840	1051	4011	437	408				7828	1753	6686	729		0 16	
1995		3439	2697	1017	3853	434	184				5471	1696	6422	723		0 35	
1996		1809	3600	477	4666	500	120		16122		7370	797	7816	833			2 4532
1997	11491	201	3457	292	3529	462	58				7049	487	5882	770			35 3531
1998		1183	3578	328	5121	1127	58	0			7347	555	8536	1878			35 1068
1999		708	3194	1868	5401	1429	0	-			6536	3098	9002	2382		0 71	
2000		429	1116	602	7399	633	. 0	-			2284	1004	14050	1055		0 82	
2001	8050	185	2632	266	3225	728		0			5392	443	5374	1213		0 77	
2002		31	3189	689	4333	1448	0				6530	1149	7222	2414		0 167	
2002		0		721	3566	1512	0				6538	1201	5944	2520		0 97	
2004		107	6526	284	4889	1639	0				13104	474	8149	2731		0 115	
2005		0		794	3353	1508	0				7485	1323	5588	2513		0 90	
2006		0		1800	2944	1455	0				7584	2999	4907	2426		0 111	
2007		0		1710	1830	1024	0	-			7631	2850	3051	1707		0 86	
2008		0		2266	3144	1401	0				11261	3776	5240	2336		0 117	
2009		0		903	1907	1056	0	-			7306	1505	3178	1759		0 114	
2010		0		993	1675	1081	0	-			9746	1655	2792	1802		0 105	
2011		0		1365	4441	1694	0	-			10386	2276	7402	2824		0 120	
2012		0		584	3550	1228	0	-			7332	973	5916	2047		0 123	
2012		88	3185	411	2466	1401	0				6461	685	4111	2335		0 117	
2014		0		676	2412	2341	0				8003	1127	4020	3901		0 124	
2015		0		1396	4254	1359	0	0			11853	2295	7089	2266	0	0 124	

Appendix 4.ix. (Continued). Input data for large salmon returns to Québec by category of data used in the run-reconstruction.

	Large retur	ns							Large return	ns								
	Minimum								Maximum									
Year	C							Other rivers		C2	C3	C4	C5	C6		Other rivers		Commercia
													[] QCLgC5_H[]					QCCmLg[]
1970		0	-			-				'	,	0			0	, ,		
1971		0	-									-) (С	
1972		0	-	-						`			-) (C	
1973		0				-		-		'		0	-) (, ,		
1974		0		-				-		`	,	0) (C	
1975		0	-	-				-		'		0	-) (
1976		0	-	-						,		9	-	-) (C	
1977		0	-	-				-					-) (
1978		0		-			,			`		0		-) (C	
1979		0					,					0	,) (C	
1980		0		-			,	, .				0	-	-) (C	
1981		0		-			, ,					0	-) (C	
1982		0					,					0) (C	
1983		0 0						-				-	-) (, ,		
1984																	4530	
1985										-								
1986													61 979					
1987													58 10558					
1988								_					76 1131					
1989													953°					
1990																		
1991																		
1992													22 4478					
1993																		
1994									_				475					
1995																		
1996								4	24117									
1997								9	19154				53 4775					
1998				347				0					92 4649					
1999					3870													
2000				560				-					33 10700					
2001								0					02 6647			, ,		
2002											927		3505					
2003											11779		19 8148					
2004													95 7387					
2005								, ,			651					, ,		
2006					3945						6904							
2007					3171			-			7406					, ,		
2008			0.0.								1059							
2009											7589							
2010								-			915							
2011	2697			1571	6007	483	S C	0	28373		9529	9 261	1001	1 805	5 0			6
2012	1791	В	3665	904	4488	313	g C	0	18837		743	4 150	748	1 522	2 0	0	3023	8
2013				989	3938	339	C	0	23135	242	846	1 164	18 6563	3 565	5 0		2895	5
2014	1095			695	1593				11504	. (4869	9 115	59 265	5 1725			2908	3 (
2015	1808	5 0	3995	1812	5048	568	C	0	19007	′ (8104	4 296	8414	4 946	6 0	0	2765	5 (

Appendix 4.ix. (Continued). Input data for small salmon spawners to Québec by category of data used in the run-reconstruction.

	Small spawr	ners					Small spawn	ers				
Year	Minimum C1	C2	C3	C4	C5	C6	Maximum C1	C2	C3	C4	C5	C6
Bugs labels												QCSSmC6 H
1970					QCSSIIICS_L							
1970							-					
1971												
1972												
1974												
1975												
1976												
1977									_			
1978							_		_			
1979					0				0			
1980		-			0			0				
1981					0			A 0				
1982	_				0					C		-
1983		-		-	0			0		C	-	-
1984					1264	5160	4			747		-
1985					4241	4384				351		
1986					5151	5133				107		
1987					6411					510		
1988					6432					731		
1989					5149					475		
1990					5437	2976				1068		
1991					3827	2001				551		
1992						3462				657		
1993					3339	1447				1435		
1994				894	3089					1596		
1995				877	2956			2923		1556		
1996	10636	1330	2260	372	3678	500	11748	1444	6030	692	6828	833
1997	8238	142	2250	266	3074	462	8836	178	5842	461	5426	770
1998	7734	995			4229	1124	8298	1218	6116	516	7643	1875
1999	8155	509			4581	1426	8834	542	5837	2883	8182	2379
2000	8291	372	693	519	5900	583	9040	401	1861	921	12551	1005
2001	5329	143			2579	658	5867	186	4140	440	4729	1137
2002	9296	31	2231	658	3405	1448	10191	36	5572	1118	6294	2414
2003	8180	0	2269	661	2826	1509	8721	0	5604	1141	5204	2517
2004	9030	29	5574	278	3962	1639	9460	49	12152	468	7222	2731
2005	6339	0	3025	716	2709	1506	6756	0	6821	1245	4945	2511
2006	8628	0	3159	1691	2372	1455	9235	0	7007	2890	4335	2426
2007	5768	0	3226	1511	1501	1024	6217	0	7099	2651	2722	1707
2008	10562	2 0	4882	1756	2522	1401	11467	0	10601	3266	4618	2336
2009	6293	0	3115	764	1633	1056	6736	0	6820	1366	2904	1759
2010	8860	0	4289	914	1311	1080	9509	0	9234	1576	2428	1801
2011	12143	0	4496	1116	3674	1688	13049	0	9762	2027	6635	2818
2012	6620	0	3152	472	2924	1225	7092	. 0	6869	861	5290	2044
2013	4959	88	2840	328	2131	1401	5314	104	6116	602	3776	2335
2014	6579	0	3239	560	2127	2175	7033	0	7297	1011	3735	3735
2015	10718	0	4709	1109	3457	1352	11518	0	10719	2008	6292	2259

Appendix 4.ix. (Continued). Input data for large salmon spawners to Québec by category of data used in the run-reconstruction.

	Large spaw	ners					Large spaw	ners				
	Minimum						Maximum					
Year	C1	C2	C3	C4	C5	C6	C1	C2	C3	C4	C5	C6
Bugs labels	QCSLgC1_L[QCSLgC2_L[QCSLgC3_L[QCSLgC4_L[QCSLgC5_L[QCSLgC6_L[QCSLgC1_H	QCSLgC2_H	QCSLgC3_H	QCSLgC4_H	QCSLgC5_H	QCSLgC6_H
1970	0	0	0	0	0	0	C	0	0	0	0	0
1971	0	0	0	0	0	0	C	0	0	0	0	0
1972	0	0	0	0	0	0	C	0	0	0	0	0
1973	0	0	0	0	0	0	C	0	0	0	0	0
1974	0	0	0	0	0	0	C	0	0	0	0	0
1975	0	0	0	0	0	0	9	0	0	0	0	0
1976	0	0	0	0	0	0	C	0	0	0	0	0
1977	0	0	0	0	0	0	C	0	0	0	0	0
1978	0	0	0	0	0	0	9	0	0	0	0	0
1979	0	0	0	0	0	0		. 0	0	0	0	0
1980	0	0	0	0	0		C	0	0	0	0	0
1981	0	0	0	0	0	0	C	0	0	0	0	0
1982	0	0						0				
1983	0											
1984	10421	7648	-				11933	-	-			
1985	9985					3351						
1986	13659											
1987	13432											
1988	15535											
1989	14645									440		
1990	12398					_						-
1991	14061	5019				357						
1992	12850											
1993	9848											
1994	10468											
1995	16562			781		154						
1996	16431	4417	-									
1997	13433					138						
1998	10402											
1999	14169											
2000	11937				5096	363						
2000	14527					348						
2001	10843					344						
2002	18832											
2003	15558											
2004	16485											
2005	14977					403						
2007	12470					303						
2007	13725					390						
2008	16489					275						
2009	19170					338						
2011	24130 16098					479 313						
2012											6678	
2013	19804					338				1579		
2014 2015	10089 16672					740 557				1128 2902		

Appendix 4.ix. (Continued). Year specific returns and spawner data for Québec for years when category splits are not available (1970 to 1983) used in the run-reconstruction.

	rnese data are	з эресніс ю тп	e 1970 to 1983	perioa wrien a	etaneu returns	by river catego	ory are nocava	iiaDie.
	Small returns		Large returns		Small spawne	ers	Large spawne	ers
Year	Min	Max	Min	Max	Min	Max	Min	Max
Bugs labels	QCSm_L[]	QCSm_H[]	QCLg_L[]	QCLg_H[]	QCSSm_L[]	QCSSm_H[]	QCSLg_L[]	QCSLg_H[]
1970	18904	28356	82680	124020	11045	16568	31292	46937
1971	14969	22453	47354	71031	9338	14007	16194	24292
1972	12470	18704	61773	92660	8213	12320	31727	47590
1973	16585	24877	68171	102256	10987	16480	32279	48419
1974	16791	25186	91455	137182	10067	15100	39256	58884
1975	18071	27106	77664	116497	11606	17409	32627	48940
1976	19959	29938	77212	115818	12979	19469	31032	4654
1977	18190	27285	91017	136525	12004	18006	44660	66990
1978	16971	25456	81953	122930	11447	17170	40944	61416
1979	21683		45197	67796		23795	17543	2631
1980		44686		161192		31226	48758	7313
1981	41667	62501	84428				35798	5369
1982			74870			25316	36290	5443
1983		26981	61488				23710	3556
1984							0	
1985							0	
1986							0	
1987							0	
1988				0			0	
1989					0		0	
1990					0	0	0	
1991	0	-		0	0		0	
1992		0			0		0	
1992				0	0		0	
1994		0		0	0		0	
1994			0	0	0		0	
1995				0	0		0	
1996	_			0	0		0	
1997				0	0		0	
					0		0	
1999		0		0				
2000		0	0				0	
2001			0	0	0		0	
2002			0		0	0	0	(
2003			0		0		0	
2004		0	0		0		0	
2005		0					0	
2006		0		0	0	0	0	
2007				0	0	0	0	
2008				0	0		0	
2009		0			0		0	
2010		0			0		0	
2011	0						0	
2012							0	
2013		-			0		0	
2014	0	0	0	0	0	0	0	
2015	0	0	0	0	0	0	0	

Appendix 4.x. Input data for 2SW salmon returns to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

	Returns of 2S	W											
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return													Point
to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	estimate
										SF19_21R2_H			
Winbugs labels	SF15R2_L[]			- 13	- 1.1	SF17R2_H[]	SF18R2_L[]				13	SF23R2_H[]	USAR2[]
1970	8243			45798		60	4744						0
1971	3587												653
1972	4980												1383
1973	6211							_					1427
1974	7264												1394
1975	4353												2331
1976	7293												1317
1977	9174			73330									1998
1978	5458				0		_						4208
1979	1472			9306									1942
1980	7102	14045	35163	48457	2			5177	14094	17318	19056	26865	5796
1981	4572	7357	11144	19268	40	77	2490	3769	8662	11471	11026	15267	5601
1982	4314	6313	21442	41643			4135	5901	4458			13871	6056
1983	3453	5280	16349	28419	17	32	3733	5241	4134	5356	9662	13836	2155
1984	3329	6092	12216	31455	13	26	2391	3573	1758	2854	15706	22627	3222
1985	4805	9500	14614	37625	8	15	921	4481	6894	12124	16541	23828	5529
1986	7831	15403	21617	55640	5	11	2274	11479	6755	11878	9891	14261	6176
1987	4836	9123	12524	32224	66	128	2611	10422	3748	6591	6922	10043	3081
1988	7152	13998	14384	36938	4 96	185	2533	10205	4393	7735	4716	6697	3286
1989	4390	8492	9113	23385			2108	8600	4808	8469	6560	9437	3197
1990	4326	8369	14269			545			3591	6320	5486	7918	5051
1991	2387	4668	14685	37736	188	361	2350	9628	2960	5213	7337	10563	2647
1992	4002			30728									2459
1993	1395												2231
1994	3960				169		1981						1346
1995	2713												1748
1996	3917						3247						2407
1997	2488						3421						1611
1998	1687						2055						1526
1999	1780												1168
2000	2270												533
2001	3779			19289			1689		1822				788
2002	2335		_				1228						504
2003	3947					557	2380						1192
2004	3005						2639						1283
2005	3422						2217						984
2006	2551	4973					2114						1023
2007	4267			15183									954
2008	2848			11066			2020						
2009	3948		10580						827				2069
2010	2978		7804		439		2049						1078
2010	7265						3633						3045
2012	3230						831						879
2012	5324						1871						525
2013	2704						1266						334
2014	3292												761
2015	3292	0462	9092	16220	/99	1195	1880	9068	335	460	248	314	/61

Appendix 4.x. (continued). Input data for large salmon returns to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

		flarge sal											
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return													Point
to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	estimate
	SF15Lg_L	SF15Lg_	SF16Lg_L	SF16Lg_	SF17Lg_L	SF17Lg_	SF18Lg_L	SF18Lg_	SF19_21L	SF19_21L	SF23Lg_L	SF23Lg_	
Vinbugs labels	[]	H[]	[]	H[]	[]					g_H[]		H[]	USALg[
1970	12681	16270	46462	49599	31	60	6161	7858	7273	9671	9691	13945	(
1971	5518	7102	28365	33409	29	29	2456	3198	5350	6773	8056	11573	653
1972	8441	16536	30146	45087	402	402	6095	6924	7460	9082	8890	12536	1383
1973	8393	16229	27771	42276	206	206	5376	6299	8049	10069	4760	6638	1427
1974	9950	19959	43249	66179	386	386	7119	7963	13138	15363	12187	16444	1394
1975	5510	10028	29826	45305	345	345	4483	4989	12261	13797	14829	20351	233
1976	9596	18969	23943	36016	575	578	3578	4223	8873	10416	16128	22175	1317
1977	11053	21779	52673	77434	606	606	5175	6280	14119	16690	19165	26183	1998
1978	7277	14332	22653	30245	0	0	5954	7201	10471	12378	9335	12342	4208
1979	2886	4971	9435	13507	459	463	1676	2315	5180	6684	5856	7903	1942
1980	8768	17340	37014	51008	2	5	4846	5951	16388	20137	21464	29480	5796
1981	9729	15652	16708	28887	40	77	3234	4332	11706	15501	12481	16743	5601
1982	7311	10700	26504	51475			5370					15303	
1983	5852			35304							10908	15235	
1984	4214	7711	12941	33321	13	26	3105	4107	2408	3909	17706	24992	
1985	7627	15080	16798	43247								26289	
1986	10305		25342	65228								15761	6176
1987	7556		15734	40483								11116	
1988	9933		17627	45267		_				12891	5360	7312	
1989	7701	14898		35812								10380	
1990	6362		23164	59479						10897		8710	
1991	4773						3052					11659	
1992	7411	14420	34573	49686	95	183	3083	11008	3657	6437		10726	2459
1993	3487		22602									5980	
1994	6600			32992							3659	4155	
1995	4171	8199			384							4289	
1996	6026	_		28035			4217			6283		6365	
1997	3828		14711	24521	387	581	4443			2726		3678	
1998	2595	_	_	26060								2437	
1999	2738											3090	
2000	3493			24778								1430	
2001	5815			29875								2501	
2002	3592		11094	18077						1029		752	
2003	6072			28749		557				2682		1289	
2004	4623		18589	32435		550						1698	
2005	5265		17008	30057	373							1121	
2006	3924		18805	32890							997	1276	
2007	6565		16018	25734						959		841	958
2008	4382		10377	19761	429					2650		1105	
2009	6074			27543								2158	
2009	4581	8972		22708						1416	1117	1398	
		22223	24960	57144								3421	
2011	11177											422	
2012	4969			21661	653								
2013	8190		10860	22992								660	
2014	4160			16849		735						299	
2015	5064	9942		23686	564	846	2442	10423	375	515	257	326	

Appendix 4.x. (continued). Input data for small salmon returns to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

	Returns o	f small sal	mon										
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return													Point
to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	estimate
	SF15Sm_	SF15Sm_	SF16Sm_	SF16Sm_	SF17Sm_	SF17Sm_	SF18Sm_	SF18Sm_	SF19_21Sm_	SF19_21Sm	SF23Sm_L[SF23Sm_H[
Winbugs labels	L[]	H[]	L[]	H[]	L[]	H[]	L[]	H[]	ᄕ	_H[]]	j	USASm[]
1970	2834	6279	47779	67697	0		264	1073	16177	24106	5306	7521	0
1971	2113	4681	38388	54120	0	0	65	265	11911	18004	3248	4541	32
1972	2185	4699	48886	69270	0	0	131	530	11587	17992	1831	2506	18
1973	3010	6668	47190	66835	5	9	516	2095	14169	22159	5474	7012	23
1974	2226	4895	78091	110470	0	0	187	757	25032	39058	10195	12901	55
1975	2393	5298	69993	98443	0	0	112	454	10860	15753	18022	23101	84
1976	8667	14696	96504	136107	14	28	299	1212	21071	33009	22835	28864	186
1977	6085	12084	30621	42689	0	0	215	871	24599	37314	13738	16671	75
1978	4350	7749	29783	39927	0	0	78	316	7621	10023	6271	7695	155
1979	4378	9495	50667	70714	2	5	1857	7536	24298	37514	15356	20517	250
1980	7994	15278	41687	58839	12	23	520	2108	34377	50250	25139	31483	818
1981	9380	17119	63278	108226	259	498	2797		31204	48945	16826	21803	1130
1982	6541	13383	78072		175				17619			15636	
1983	2723	4638	24585	41332	17	32	212		9313	14068	9270	12592	295
1984	12003	15867	28714		17	32			18382	29867			
1985	7003	15516	53393	92224	113	217	730	3167	24384	39541	13056	17928	392
1986	10813	23926	103230	178295	566		965	3854	24369	39663		20183	
1987	9630	21220	74485	128644	1141	2194	1646		27269	44266	13358	17662	1128
1988	13168	29092	107071			2963	1381		24509	39750		23084	
1989	6357	13900	66069		400					41557	17579		
1990	7880	17314	73020		1842					48039			
1991	4441	9828	53453			3028	1160			15955		17685	
1992	8853	19614	142416	204708			994	3531	13754	22269	13563	18404	1194
1993	5783	12812	70090				1146			21681	7610		
1994	9136	20208	41773				671			5393			
1995	2902	6429	44357		658		543			13873			
1996	6034	13370	32067		710	1065	2431		13120	22293		15256	
1997	5797	12845	14377		517		561	2134					
1998	6288	13932	21965		508		633						403
1999	4936	10929	21494		413		705		3971	5337			
2000	7459	16520	31923										
2001	4947	10953	26496		415		822						
2002	11719	25958	40432		390		844					3991	450
2003	3119	6904	26530	39772	515	773	773			3837			
2004	12091	26783	43242				1092			5192			
2005	4117	9116	28441		343						3597		
2006	8724	19322	30671		331		869						
2007	4259	9430	23038				718						
2008	13601	30129	25722		298		1245		7282		5924		814
2009	5169	11445	10819				302		2066				
2010	8187	18132	48123		258		877						
2011	10234	22668	39511										
2012	4350	9631	6914		291	436	211		346				
2012	4661	10320	10106				429						78
2014	3697	8183	6528		222		219					910	
2014	6228	13793	25575							2648		2131	

Appendix 4.x. (continued). Input data for 2SW salmon spawners to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

	Spawners	s of 2SW											
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return													Point
to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	estimate
	SF15S2_	SF15S2_	SF16S2_	SF16S2_	SF17S2_	SF17S2_	SF18S2_	SF18S2_	SF19_21	SF19_21	SF23S2_	SF23S2_	
Vinbugs labels	L[]	H[]	L[]	H[]	L[]	H[]	L[]			S2_H[]	L[]		USAS2[
1970	1156	3252	5346	8242	18	47	304	1587	2388	4234	1536	4846	(
1971	510	1434	6724	11354	0	0	133	694	1418	2513	3612	6576	490
1972	2367	6656	17031	31450	0	0	148	775	1616	2865	6472	9806	1038
1973	2873	8081	19277	33170	0	0	165			3984	2752	4412	1100
1974	3620	10183	31192	52012	0	0	151	790	2878	5103	8123	12046	1147
1975	1769	4975	18536	31972	0	0	91	473		3523	10987	16209	1942
1976	3530	9928	11842	22152	1	4	116	604	1935	3432	10071	15583	1126
1977	4412	12408	30623	54071	0	0	198	1033	2559	4539	12013	18568	643
1978	2622	7375	6998	13535	0	0	223	1166	1948	3455	5346	8076	3314
1979	527	1482	3000	5806	3	7	115	598	1419	2517	3772	5650	1509
1980	3440	9677	17667	30961	1	4	198	1033	4170	7394	12023	19005	4263
1981	1380	3880	2392	10515	36	73	196	1027	3631	6439	3642	7014	4334
1982	991	2786	8418	28619	8	23	253	1322	1158	2053	4475	7939	4643
1983	906	2547	5516	17586	15	30	210	1100	1579	2800	468	3561	1769
1984	2656	5402	11650	30889	13	26	259	1148	1416	2512	12280	18798	2547
1985	4514	9180	14019	37030			871	4359	6761	11990	11885	18624	4884
1986	7279	14804	20606	54630	5	11	2164	11213	6624	11748	7224		5570
1987	4122	8383	11414	31114	66	128			3676	6519	5628	8597	278
1988	6582	13386	13801	36355	96	185	2451	9954	4322	7664	3420	5248	3038
1989	3944		8466	22739			2042						
1990	3886	7903	13669	36039	284		1829	7491			4926	7292	
1991	2193		14200	37251	188		2275				6080		2416
1992	3639		20770			183		9324			5826		
1993	1239		15239	59907	22						3291	3654	
1994	3639		13418	24653			1920				2387	2680	1344
1995	2519		25326	_	380								
1996	3688	7502	10743	18662	388		3166				4009		2407
1997	2316		8106	13754	385		3334						1611
1998	1512	_		10548	382		2000				1068		
1999	1581		6589		379						1934		1168
2000	2057			11408	376						805		
2001	3521	7161	13688	18674	374		1654				1699		149
2002	2120			8808		557	1203				317		511
2003	3683	_	10593	16372			2333				878		1192
2004	2770		10144	17965	365			11219		1401	1238		1283
2005	3175		10755	19354	371	560					726		
2006	2329	4737	9336	16594	390	587	2062				796		1419
2007	3994			14644						809	530		
2008	2618		5376	10584	429			9180			736		2809
2009	3684		10062	16500	401	602		7122		1118		1774	
2010	2743		7335	11078							726		1482
2010	6902		20445	47555	652					2023			
2011	2988		7603	14713						828			
2012	5019		6909	15204	989					2834			
2013	2478		5606		709								566
2014	3048		9518	11217 15988		1063					163 244		1509

Appendix 4.x. (continued). Input data for large salmon spawners to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

	Spawners of	large salmon											
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return													Point
to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	estimate
									SF19_21SLg_				
Winbugs labels		SF15SLg_H[]				0- 13		SF18SLg_H[]				SF23SLg_H[]	USASLg[]
1970	1779			8926	18		395					5705	
1971	785		7324	12369	0					3264			
1972	4011	11282			0			_					
1973	3883												1100
1974	4960	13949											1147
1975	2239				0			_	_				1942
1976	4644	13063		25940		4	-						1126
1977	5315			57097	0								
1978	3496	9833		15720									3314
1979	1033	2906	4355	8426	3	7	149	688	1944	3448	4217	6264	1509
1980	4248	11947	18597	32590	1	4	257	1187	4849	8598	13190	21206	4263
1981	2935	8256	3586	15765	36	73		1181	4907	8702	3794	8056	4334
1982	1679	4723	10405	35376					2464	4369	4903	9059	4643
1983	1535	4317	6852	21846			273		2506	4445	92	4419	1769
1984	3362	6838	12341	32721	13	26	337	1320	1940	3441	13675	20961	2547
1985	7164	14571	16114	42563	8	15	1131	5010	8347	14803	13104	20811	4884
1986	9577	19479	24157	64044	5	- 11	2811	12889	10515	18647	8004	12623	5570
1987	6441	13099	14340	39088	66	128	3291	11711	5835	10347	6343	9594	2781
1988	9141	18592	16913	44553	9 6	185	3183	11442	7203	12773	3835	5787	3038
1989	6919	14072	12965	34822	149	. 287	2652	9651	6862	12168	7099	10086	2800
1990	5715	11623	22190	58504	284	545	2376	8611	6087	10793	5576	8051	4356
1991	4386	8920	23472	61572	188	361	2955	10803	4045	7173	6833	10180	2416
1992	6738	13704		48697	95				3594				
1993	3099				22								
1994	6065			32682	166		2493						1344
1995	3873			43778									1748
1996	5674		15755		388		4112						2407
1997	3563	7247	13955		385		4330						1611
1998	2326						2597						1526
1999	2433											2601	1168
2000	3165		15321	24069	376								
2001	5417	11018		28923			2148						
2002	3261	6633				557	1562			1021	442		
2003	5666						3029			2661	919		1192
2004	4261	8666		31517			3351						1283
2005	4884	9934				560	2807						
2006	3583			31911	390		2678						
2006	6145			24820	409								
2007	4028			18901	409		2547						2231
2009	5668			26612		602							2318
2009	4221	8584	14382	21722									
2011	10619		24053	55948									
2012	3230	6338		21019						1048			2054
2013	7721	15704											
2014 2015	3812 4689		8318 13899	16642 23347	489 562		1616 2393						572 1519

Appendix 4.x. (continued). Input data for small salmon spawners to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

	Spawners of	small salmon											
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return to rivers	Min	Max	Min		B4:	Max	Min		Min	Max	Min	Max	Point
to rivers	IVIIII	Wax	IVIII	Max	Min	Wax	IVIII	Max		SF19 21SSm		IVIAX	estimate
Winbugs labels	9E1599m I II	SE15SSm HII	SE16SSm []	SE16SSm HII	SE17SSm []	SE17SSm HII	SE18SSm []	SF18SSm_H[]			SF23SSm_L[]	SESSES HII	USASSm[]
1970	1417												03A3311[]
1970						-							
	1056			38195							1216		29
1972	1034			48023									17
1973	1505												
1974	1098							_					40
1975	1195			78888					5819				
1976	2480			104130					14196				151
1977	2467			25338					15120				
1978	1398			24833				248		5259			127
1979	2104			51876									247
1980	2996			44943									722
1981	3183			80370									
1982	3038	9027	51324	106423				6847	11244				290
1983	820	2486	13298	30045	10			674	5653	10408	5132	8454	255
1984	1620	4971	7389	28271	10	25	177	1200	13658	25143	10290	16412	540
1985	3557	10936	32275	71106	66	170	145	1788	18024	33181	8164	13036	363
1986	5589	16990	71918	146983	330	852	63	1729	18187	33481	10725	16634	660
1987	4867	14920	49971	104131	665	1718	527	3075	20213	37210	10257	14561	1087
1988	6664	20468	71967	149800	899	2320	344	2388	18125	33366	13061	19764	923
1989	3191	9741	37696	85724	233	603	232	1650	18973	34928	13124	20066	1080
1990	3996	12190	46902	99996			229	1750	22080	40648	10025	15381	617
1991	2215	6872	39648	78522	919	2371	271	2068	7363	13556	9495	14139	
1992	4426			178949									
1993	2891			1570 56									444
1994	4554			43764	118					4900			427
1995	1451		34650	53746									
1996	3017			29260									
1997	2899		8702										
1998	3144			21387									403
1999	2465			17943					3895		5196		419
2000	3727			26196									270
2001	2470			22815									266
2001	5857			35509					5180				
2002	1557			24997					2829				
2003	6043			40613					3833				319
2005	2056								2854 5119				319
2006	4359			35085									450
2007	2127			29105									297
2008	6798			30904									814
2009	2581			13313						2773			241
2010	4090			44243									525
2011	5114			45125						4864		5867	1080
2012	2172			7895									
2013	2328			12437									78
2014	1845			7361	220								110
2015	6036	13539	24339	33200	229	343	394	4151	1960	2647	1654	2108	150

Appendix 5: Model Walkthroughs

Summaries of the data preparation, model running and output processing were presented at a one-day workshop prior to the 2014 WGNAS meeting and provided step by step walkthroughs of the assessment processes. Where appropriate these have been updated in 2015.

NEAC pre-fishery abundance and national conservation limit model in R

[NB: Instructions apply to model version: "NEAC_PFA_CL_RR_model_2015-varM_v12" as used in 2015]

1) Introduction

This program performs the run-reconstruction estimation of pre-fishery abundance (PFA) of maturing and non-maturing 1SW salmon for each country (and region) in the NASCO-NEAC area. PFA is estimated for January 1st in the first sea winter. The program also establishes the pseudo stock–recruitment (S–R) relationship between lagged egg deposition and Total 1SW PFA, and applies a hockey-stick S–R analysis to estimate the National/Regional Conservation Limit where river-specific CLs are not available. The original model is described by Potter *et al.* (2004); minor changes to the estimation approach used for different countries and regions have been reported in the annual reports of WGNAS.

The model also estimates the proportion (mean and SE) of the Faroes catch originating in different countries/regions based on the genetic analysis of scales collected in the fishery between 1993 and 1995 and the estimated PFA for each country/region since 2001, when no fishery has been operating at Faroes. This requires the model to be run once to provide the catch proportions and then run a second time to provide a full PFA assessment. The catch proportions are also used in the Catch Options model.

2) To get started

- 2.1) Load RStudio or R;
- 2.2) Set up a folder from which you will run the program;
- 2.3) Use folder and file names without spaces;
- 2.4) Put the program, the input files (annual and multiannual) and the summary data file (see 6f) in this folder.

3) Input Data

- 3.1) Annual data (filenames: Annual-data-XX-YY.txt)
 - 3.1.1) There is a file for each country (XX) and region (YY) which contains the 40+ year time-series of data on catches, exploitation rates and non-reporting rates (plus additional data for some countries).
 - 3.1.2) To read the .txt files, it is easiest to open them from within Excel. i.e.
 - Open Excel;
 - select the correct folder;
 - click on 'Open'

 You will probably need to change the setting in the lower right corner of the open box from 'Excel files' to 'All files';

- Double-click on the file you want to open and it should open the 'Text Import Wizard';
- select 'finish' (If this doesn't work reopen the file, but select 'Delimited' at step 1, 'Tab' at step 2 and 'General' at step 3.)
- 3.1.3) Do not add any formatting to the file. If loading a new version of a file that has been saved in Excel (e.g. after addition of a new year's data), re-save the file by clicking 'Save As' and selecting 'Text (Tab delimited)' from the 'Save as type' list. This will remove the formatting and add the .txt extension.
- 3.1.4) Do not change the file name.
- 3.1.5) Close and save the file before running the programme. You will be prompted to confirm that you want to lose the formatting; click 'yes'.
- 3.2) Multiannual-data (file-name: 'Multiannual-data.txt' or similar)
 - 3.2.1) This file contains most of the other parameters used in the model including: smolt age composition, fecundity and sex ratios by region, M, etc.
 - 3.2.2) The second value listed is the 'lastdatayear' which needs to be updated to the latest year for which data are provided in the Annual-data-XX files.
 - 3.2.3) The file is not formatted in columns so can be read easily in Notebook, which should be selected automatically if you click on the file to open it. (NB: If you open the file in Excel, don't save it because it will probably add " " marks to each line.
 - 3.2.4) All blank lines and lines starting with '#' are ignored in this file. Apart from these:
 - The first line must start with 'list(
 - The last line must be ')'
 - All other lines must be 'variable name' <- number, followed by a comer (except for the last data line which has no comer).
 - 3.2.5) If the module estimating the composition of the Faroes catch is run (see below) the new values must be inserted at the end of the multiannual data file in place of the current ones.
 - 3.2.6) Save the file before running the model.

4) Model structure

- 4.1) Introductory section: contains working directory, source files and various parameters controlling the way the program runs (some of these will need to be changed for your laptop (see below).
- 4.2) Functions: functions are sections of code that the program calls up to repeat the same job. They have to be run before they are first called by the program; this is achieved by placing them at the beginning of the code. The main functions run the hockey-stick analysis for the NCL model and output certain figures and tables.

4.3) Faroes and Greenland sections: these sections calculate the harvest in the distant water fisheries.

- 4.4) NEAC country/regions sections: there is a section for each country (in alphabetic order) and region to calculate the main outputs of the R-R model.
- 4.5) Output summaries: this section creates NEAC summary figures and tables and the country/region data files for the Winbugs Forecast Model.
- 4.6) Faroes catch composition: The final section estimates the proportion (mean and SE) of the Faroes catch originating in different countries/regions. This requires the model to be run once to provide the catch proportions and then run a second time to provide a full PFA assessment.

5) Running the code from RStudio

- 5.1) Open R Studio
- 5.2) Select "File/Open File" and use the browser to select and open the code file; the code should open in the Top left panel. The code will have a name like "NEAC_PFA_CL_RR_model_2015_xxx"
- 5.3) If you have been using the code recently, you can select "File/Recent Files" and select the file from the drop-down list (if it is there); you can open several code files simultaneously and they appear as tabs above the Top Left panel.
- 5.4) To set up the code for your PC/laptop, R-click on the code and scroll down to:

line 40 -enter the full path name of the working directory (replace the text between the parentheses with the full pathname of the folder containing the code on your laptop (e.g. "D:/Modelling_NEAC/PFA_NCL_R/2014").

line 45 ensure that the text between the parentheses shows the correct filename for the multi-annual data file.

lines 77–86 -select which countries you wish to run the assessment for by setting "run-XX": 1 = run country XX; 0 = do not run. The summaries will only be run if all countries are set to 1.

line 82 -set "PrintFigs" equal to '1' to output the summary figures (or any other value not to output them); otherwise "0".

line 89 -set "WinbugsFiles" equal to '1' to output the data files for the Bayesian forecast model (or any other value not to output them); otherwise "0"...

line 92 -set "PrintCountryTables" equal to '1' to output summary output data for each region that is run (or any other value not to output them); otherwise "0"...

line 98 -set "RunFaroeseCatchSplitEstimation" equal to "TRUE" to run the estimation of the Faroes catch composition; otherwise "FALSE".

5.5) You do not need to save your changes before you run the code. [If you wish to save any changes, use "File/Save" or "File/Save As" as

normal. It's a good idea to include the extension ".R". NB: You will be prompted to save the file before you close it.]

- 5.5.1) To clear the 'console' area (lower left panel) press "Ctrl-L"
- 5.5.2) To run the program press "Ctrl-Alt-R"
- 5.5.3) You will see when part of the code run in console area. Errors will show in red. The run is complete when the final line shows ">"

6) Running the program from R

- 6.1) Open R Studio
- 6.2) Select "File/Open script" and use the browser to select and open the code file; the code should open in a separate panel. The code is currently called "NEAC_PFA_CL_RR_model_2014"
- 6.3) To set up the code for your PC/laptop, R-click on the code and scroll down to:
 - # SET WORKING DIRECTORY (wd): In line starting "wd <-" replace the text between the parentheses with the full pathname of the folder containing the code on your laptop (e.g. "D:/Modelling_NEAC/PFA_NCL_R/2014").
 - # SET "run_XX": in the lines starting "run_XX <-" select which countries you wish to run the assessment for by setting "run-XX": 1 = run country XX; 0 = do not run. The summaries will only be run if all countries are set to 1.
 - # SET 'PrintFigs' set "PrintFigs" equal to '1' to output the summary figures (or any other value not to output them).
 - # SET 'WinbugsFiles: set "WinbugsFiles" equal to '1' to output the data files for the Bayesian forecast model (or any other value not to output them).
 - # SET 'PrintCountryTables': set "PrintCountryTables" equal to '1' to output summary output data for each region that is run (or any other value not to output them).
 - "RunFaroeseCatchSplitEstimation": set "RunFaroeseCatchSplitEstimation" equal to "TRUE" to run the estimation of the Faroes catch composition; otherwise "FALSE" [SEE BELOW]
- 6.4) You do not need to save your changes before you run the code, but you may wish to save a version to be safe. To do this use "File/Save" or "File/Save As" as normal. It's a good idea to include the extension ".R". NB: You will be prompted to save the file before you close it.
- 6.5) To run the program select "Edit/run all"
- 6.6) You will see when the code runs in the 'R console' panel. Errors will show in red. The run is complete when the final line shows ">"
- 7) Running the Faroes stock composition
 - 7.1) The 'Multiannual-data' file contains the latest estimates of the composition of the Faroes catch by European country/regions based

on the results of the genetics analysis reported in the 2015 WGNAS report and the 2001-14 PFA outputs. These estimates may be updated if new genetics data are provided or additional years of PFA estimates are to be included.

- 7.2) To run the estimation, ensure that all "Annual-data-XX" files have been updated.
- 7.3) SET 'RunFaroeseCatchSplitEstimation' to "TRUE" and run the model.
- 7.4) The new stock composition parameters will be output in the file "Faroes_split_estimate.txt"; these data should be copied into the end of the "Multiannual-data" file to replace the values already there. These data are also required for the Catch Options models.
- 7.5) Reset SET 'RunFaroeseCatchSplitEstimation' to "FALSE" and run the model again to produce full updates of the PFA estimates.
- 8) Output files

The program produces the following outputs (if requested):

8.1) National plots: (filenames "Figure-XX")

PDF files showing the national plots currently used in the WG report. This includes: maturing and non-maturing 1SW PFA; returns and spawners for 1SW and MSW; homewater exploitation rates; and total catches (incl. non-reported) for each country (XX). It also shows the pseudo stock-recruitment hockey-stock plots for each region; these show the estimated CL, where this is used in the assessment.

8.2) Regional data: (filenames "Region_data_XX_YY")

Excel files showing PFA, returns, catch, exploitation rates and spawners for 1SW and MSW fish and total eggs and lagged egg estimates for each country (XX) and region (YY).

8.3) Input files for Forecast analysis: (filenames: "Winbugs_Data_XX_YY")

Excel files for each country/region containing mean and s.d. estimates for the simulations for lagged eggs, 1SW returns and MSW returns.

- 8.4) Summary tables by country:
- Median spawner numbers
- Conservation limits and SERs
- Maturing 1SW PFA
- 1SW returns
- 1SW spawners
- Non-maturing 1SW PFA
- MSW returns
- MSW spawners
- 8.5) Summary plot for N-NEAC and S-NEAC
- 8.6) A formatted Excel workbook is set up to link to the output files and format the tables ready for use in the WGNAS report.
- 9) Common problems

9.1) The code will crash if an output file (Figure or Table) is left open. The error message (in red) may say:

• Error: cannot open file 'Fig-XX'

or

- Error in: : cannot open file 'Region_data_XX.csv': Permission denied
- 9.2) It doesn't matter if an input file is open, but the program may not read the latest version if it has not been saved.
- 9.3) More problems to be added when they are found!

