Stock Annex: Atlantic salmon (Salmo salar)

Stock specific documentation of standard assessment procedures used by ICES.

Stock	Atlantic salmon
Working Group	Working Group on North Atlantic Salmon (WGNAS)
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1. General

1.1. Stock definition

1.1.1. Background

Atlantic salmon, *Salmo salar* L., have a wide range of life-history strategies. Most forms are anadromous, however, with a juvenile phase in freshwater followed by a period at sea feeding and growing, during which the fish undergo extensive migrations in the open ocean, before they migrate back to freshwater to breed. Most Atlantic salmon return to their river of origin to spawn. This precise homing behaviour has resulted in groups of fish originating in different rivers or tributaries becoming genetically distinct as they adapt to the particular conditions that they face in their home river and along their migration routes. As a result, fish from one river or tributary can differ from fish originating in other rivers/ tributaries which have become adapted to a different set of conditions. These subgroups comprise genetically distinct 'populations' and these are regarded as the basic biological units of the Atlantic salmon species.

Large rivers and their tributaries can support several, genetically distinct populations, each with separate spawning areas within the main-stem of the river or its tributaries. In most instances, however, it is not possible to demarcate clear population boundaries within a river, and managing stocks and fisheries at this level of detail would be very complex. Thus, while there is a need to protect the sustainability of these units, the primary management unit (e.g. for reporting catch statistics and regulating fishing) is generally taken to be the river stock, comprising all fish originating in eggs laid within the river.

Atlantic salmon are native to the temperate and Subarctic regions of the North Atlantic Ocean and there are over 2000 rivers draining into the North Atlantic that support the fish, about 1500 of which discharge into the Northeast Atlantic. In this area, salmon distribution extends from northern Portugal to northern Russia and Iceland, while in the Northwest Atlantic, the species ranges from northeastern USA (Connecticut) to northern Canada (Ungava Bay).

Ideally, the management of all individual river stocks, and the fisheries that exploit them, might be based upon the status of each individual population. This is not always practical, however, particularly where decisions relate to the management of distant water salmon fisheries, which exploit large numbers of stocks originating in broad geographic areas. WGNAS has therefore had to consider how populations or river stocks should be grouped in providing management advice. For this purpose, groups have been established which fall within the meaning of a stock as 'an exploited or managed unit' (Royce, 1984) and that are consistent with the ICES (1996) definition of salmon 'stocks' as 'units of a size (encompassing one or more populations) which provide a practical basis for the fishery manager'. The issues around the grouping of Atlantic salmon stocks for the provision of management advice are reviewed in detail in Crozier *et al.* (2003). Such stock groupings have typically been referred to as stock complexes.

Salmon mature at various sea ages, typically returning to freshwater to spawn after one to three years at sea, but also sometimes at older sea ages; this varies widely between populations. Those salmon that return after one year at sea are referred to as one-sea-winter (1SW) salmon, or grilse, with older fish categorised as 2SW, 3SW, etc. In practice, however, for management purposes these older sea age fish are typically aggregated and collectively referred to as multi-sea-winter (MSW) salmon. The sea age when salmon become sexually mature depends on genetics as well as growing conditions in the sea, and possibly freshwater, although the precise proximate factors initiating homeward migration are unknown (Hansen and Quinn, 1998). The sea age of Atlantic salmon is important in the context of stock definition since these different groups of fish have different migration routes, return at different times and are differentially exploited in fisheries. Thus, for example, it is only potential MSW salmon that are exploited in the distant water salmon fishery that operates off the west coast of Greenland.

1.1.2. Management framework for salmon in the North Atlantic

The advice generated by WGNAS is in response to Terms of Reference posed by the North Atlantic Salmon Conservation Organisation (NASCO), pursuant to its role in international management of salmon. NASCO was set up in 1984 by international convention (the Convention for the Conservation of Salmon in the North Atlantic Ocean), with a responsibility for the conservation, restoration, enhancement, and rational management of wild salmon in the North Atlantic. NASCO now has six Parties that are signatories to the Convention, Canada, Denmark (in respect of Faroe Islands and Greenland), the EU (which represents its Member States), Norway, Russia and the USA. While sovereign states retain their role in the regulation of salmon fisheries for salmon originating in their own rivers, fisheries within the jurisdiction of one Party that exploit salmon originating in the rivers of another Party may be regulated by NASCO under the terms of the Convention. This is currently the case for the distant water salmon fisheries at Greenland and Faroes.

NASCO discharges these responsibilities via three Commission areas shown below:



While homewater fisheries are not regulated directly by NASCO, national/ regional jurisdictions seek to comply with NASCO agreements and guidelines in exercising their responsibilities. In particular, NASCO's Agreement on the Adoption of a Precautionary Approach states that an objective for the management of salmon fisheries is to maintain the diversity and abundance of salmon stocks, and NASCO's Standing Committee on the Precautionary Approach interpreted this as being "to maintain both the productive capacity and diversity of salmon stocks" (NASCO, 1998).

NASCO's Action Plan for Application of the Precautionary Approach (NASCO, 1999) provides an interpretation of how this is to be achieved:

- "Management measures should be aimed at maintaining all stocks above their conservation limits by the use of management targets".
- "Socio-economic factors could be taken into account in applying the precautionary approach to fisheries management issues".
- "The precautionary approach is an integrated approach that requires, *inter alia*, that stock rebuilding programmes (including as appropriate, habitat improvements, stock enhancement, and fishery management actions) be developed for stocks that are below conservation limits".

In requesting scientific advice from ICES, NASCO asks for an annual review of events in the salmon fisheries and of the status of salmon stocks around the North Atlantic; NASCO also requests management advice for stocks in each of the Commission Areas. In fulfilling these requirements, three specific purposes have been identified for which stock groupings may be required (Crozier *et al.*, 2003):

- providing descriptions of the status of stocks;
- developing models to estimate and/or forecast pre-fishery abundance (PFA); and
- developing management advice for the distant water fisheries.

Crozier *et al.* (2003) further noted that there is no reason to assume that the same stock groupings should be used for all these purposes, indeed both the criteria used (e.g. geographical or biological features) and the resulting groups are likely to differ.

1.1.3. Stock groupings used by WGNAS in providing management advice

As noted above, Atlantic salmon would, ideally, be assessed and managed on the basis of river-specific stock units. In reality, <25% of the rivers with salmon populations in the North Atlantic are so assessed (Chaput, 2012; ICES, 2013). Consequently, stock status is often, of necessity, assessed at broader regional, national and subcontinental scales. While there might be merit in grouping stocks according to biological criteria (which could cross jurisdictional boundaries), it has generally been considered that the difficulties of collecting data in a similar format in different jurisdictions is likely to outweigh the benefits of using such groups (Crozier *et al.*, 2003). It is also recognised that compilations of data on stocks within each jurisdiction are of importance to regional / national managers. As such, regional / national stock groups are typically used by ICES in providing advice on the status of stocks, with additional information compiled on biological groups (e.g. sea ages) as required.

ICES has previously provided information on the status of stocks in the Northeast Atlantic Commission (NEAC) area by region or by country (as well as sea age). For the North American Commission (NAC) area similar information is provided for the USA and the five main provincial regions in eastern Canada: Labrador, Newfoundland, Québec, Gulf and Scotia-Fundy.

In providing management advice for the mixed-stock distant-water fisheries, broader scale stock groupings have been considered appropriate. For the NAC area this is based on the six geographic regions of North America detailed above. For the NEAC area, the following national groupings have been used in recent years to provide NASCO with catch advice or alternative management advice for the distant-water fisheries at West Greenland and Faroes.

Southern NEAC countries	Northern NEAC countries
France	Russia
reland	Finland
K (Northern Ireland)	Norway
K (England & Wales)	Sweden
JK (Scotland)	Iceland (north/east regions) ¹

¹ The Iceland stock complex was split into two groups for stock assessment purposes in 2005 (ICES, 2005), largely on the basis of tag-recapture information. Prior to 2005, all regions of Iceland were considered part of the Northern NEAC stock complex.

These groups were deemed appropriate by WGNAS as they fulfilled an agreed set of criteria for defining stock groups for the provision of management advice that were considered in detail at the 2002 WGNAS meeting (ICES, 2002) and re-evaluated at the 2005 WGNAS meeting (ICES, 2005). ICES subsequently noted, however, that provision of catch advice for NEAC stocks in the distant water fisheries should preferably be based on a larger number of smaller management units, similar to those used in the NAC area (ICES, 2010a; 2011). Such an approach was developed at the 2013 WGNAS meeting (ICES, 2013) and indicative catch advice was provided at the coun-

try level as well as the Southern and Northern NEAC stock complexes. ICES is awaiting feedback from NASCO on the choice of management units.

Salmon from most NEAC stocks mix in the Norwegian Sea in autumn and winter, and are exploited by the fishery at Faroes. While there is evidence that some salmon from NAC rivers have been caught in the Norwegian Sea, they are currently not considered in the NEAC assessments, although this is now under review. Recent genetic information suggests that more North American fish than previously thought were exploited in the fishery at Faroes. Further details on the results of these investigations are provided in Section 3.3.3 of this report and potential options tor accounting for these fish in future catch advice is provided in Section 3.6. To date, consideration of the level of exploitation of national stocks in the Faroes fishery (ICES, 2005) resulted in the proposal that catch advice for the fishery should be based upon all NEAC area stocks and both 1SW and MSW fish.

In contrast, the fishery to the west of Greenland operates in an area where salmon from all North America and some Northeast Atlantic stocks mix in their second summer at sea. Catch advice for this fishery is thus based on non-maturing (potential MSW) fish from all regions of North America, while consideration of the level of exploitation of national stocks in the fishery from NEAC, resulted in catch advice being based upon only Southern NEAC non-maturing **1SW** (potential MSW) fish (ICES, 2005).

1.2. Fisheries

Most exploitation of Atlantic salmon is restricted to fisheries close to or within the rivers of origin of the stocks; these homewater fisheries take adult fish that are mainly returning to these rivers to spawn. As noted above, these fisheries are not directly regulated by NASCO since the Parties retain responsibility for the regulation of fisheries for salmon originating in their own rivers. However, NASCO can regulate fisheries undertaken by a Party that take salmon originating in another Party's rivers, such as is the case for the distant-water fisheries at Greenland and Faroes. These fisheries take salmon originating in a large number of rivers over a wide geographical range.

1.2.1. The Northern Norwegian Sea Fishery

A longline fishery for salmon in parts of the Norwegian Sea, north of latitude 67°N, commenced in the early 1960s. Several countries participated in this fishery and the pattern of fishing, area of operation and catches changed markedly over the years. At its peak in 1970 this fishery harvested almost 1000 tonnes of salmon.

The Convention for the Conservation of Salmon in the North Atlantic Ocean, which resulted in the formation of NASCO, came into force in October 1983. The Convention created a large protected zone, free of targeted fisheries for Atlantic salmon in most areas beyond 12 nautical miles from the coast. An immediate effect was the cessation of the salmon fishery in the Northern Norwegian Sea outside the Faroes EEZ, with the last catches in this area reported in 1984 (ICES, 2013).

In the late 1980s and early 1990s, NASCO acted through diplomatic initiatives to address fishing for salmon in international waters by vessels registered to non-NASCO Parties. There have been no reports of such activities since this period.

1.2.2. The Faroes fishery

The fishery in the Faroes area commenced in 1968 with a small number of vessels fishing up to 70 miles north of the Faroes; initially catches increased slowly up to 40 tonnes in 1977. Danish vessels participated in the fishery between 1978 and 1982 and, at the same time, catches started to increase rapidly, peaking at 1025 tonnes in 1981. Several factors contributed to this increase: the season was extended, more vessels entered the fishery, and the fishery shifted northwards.

From 1982, the Faroese Government agreed to a voluntary quota system, involving a total catch of 750 tonnes in 1982 and 625 tonnes in 1983 (255 boats allowed 25 tonnes each). Since NASCO's establishment, regulatory measures or decisions have been agreed by the Northeast Atlantic Commission in most years (Table 1.2.2.1). These have resulted in greatly reduced allowable catches in the Faroese fishery, reflecting declining abundance of the salmon stocks. There has been no commercial salmon fishery targeting salmon around the Faroes since the early 1990s. Catches in the fishery are presented in Figure 1.2.2.1.

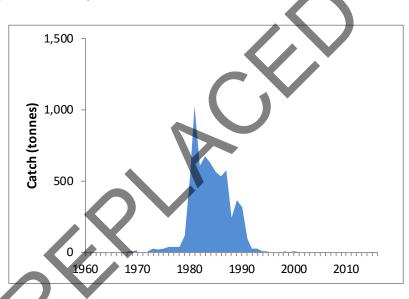


Figure 1.2.2.1. Nominal catch of salmon (tonnes, round fresh weight) in the Faroese longline fishery, 1960–2016.

Year	Allowable catch (tonnes)	Comments/other details in the measures/decisions
1984–1985	625	
1986	-	
1987–1989	1790	Catch in any year not to exceed annual average (597t) by more than 5%.
1990–1991	1100	Catch in any year not to exceed annual average (550t) by more than 15%.
1992	550	
1993	550	
1994	550	
1995	550	
1996	470	No more than 390 tonnes of the quota to be allocated if fishing licences issued.
1997	425	No more than 360 tonnes of the quota to be allocated if fishing licences issued.
1998	380	No more than 330 tonnes of the quota to be allocated if fishing licences issued.
1999	330	No more than 290 tonnes of the quota to be allocated if fishing licences issued.
2000	300	No more than 260 tonnes of the quota to be allocated if fishing licences issued.
2001–2003	No quota set	It is the intention of the Faroese authorities to manage the fishery in a precautionary manner with a view to sustainability, and to make management decisions with due consideration to the advice from ICES concerning status of stocks contributing to the fishery.
2004–2006	No quota set	It is the intention of the Faroese authorities to manage the fishery on the basis of the advice from ICES concerning status of stocks contributing to the fishery in a precautionary manner with a view to sustainability and taking into account relevant factors such as socio-economic needs and other fisheries on mixed-stocks.
2007	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.

Table 1.2.2.1. Summary of Regulatory Measures agreed by NASCO for the Faroese Salmon Fishery (courtesy of NASCO).

Year	Allowable catch (tonnes)	Comments/other details in the measures/decisions
2008	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2009	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2010	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2011	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2012	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2013–2015	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2015/2016– 2017/2018	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs

Note: The quotas for the Faroe Islands detailed above for the period 1984–2000 were agreed as part of effort limitation programmes (limiting the number of licences, season length and maximum number of boat fishing days) together with measures to minimise the capture of fish less than 60 cm in length. The measure for 1984/1985 did not set limits on the number of licences or the number of boat fishing days.

The Faroes salmon fishery operated over winter from November through to May. The salmon caught in the fishery originated almost entirely from European countries with salmon from many countries being present in the area (Jacobsen *et al.*, 2001). Small numbers of tagged fish originating in North America were also recaptured in the fishery (e.g. ICES, 1991), but excluded from catch advice. Genetic investigations, based on salmon scales removed from fish caught in the fishery in the 1980s and 1990s, suggested North American fish may make a larger contribution to the Faroes

fishery than originally indicated (ICES, 2015). There was no consistent seasonal trend in the estimated proportion of North American fish in the catches at Faroes and so the overall percentages for 1SW (5.7%) and MSW (20.5%) salmon have been used in subsequent analyses. (ICES, 2015). WGNAS has been asked to consider the implications of the findings in providing future catch advice to NASCO.

The fishery exploited mainly 2SW fish, although some 1SW and 3SW fish were also caught. Small salmon (<60 cm total length) in their first winter at sea were required to be discarded. Large numbers of farmed salmon were also observed at Faroes and there is evidence that farmed salmon escaping from net pens in Norway entered this area (Hansen *et al.*, 1987; Hansen and Jacobsen, 2003). Such farmed fish accounted for a significant proportion of the catch; in the early 1990s, the proportion of farmed fish in this area was estimated at between 25 and 40% (Hansen *et al.*, 1999).

Tagging studies (of adult fish caught in the fishery) have indicated that some fish caught at Faroes were apparently on their way westwards, as they were reported from West Greenland later the same year (Jákupsstovu, 1988). However, salmon tagged at West Greenland were also reported in the area north of the Faroes the following year (ICES, 1984). Thus, salmon of European origin are believed to move through the Faroese area on their way to the feeding areas in the West Atlantic as well as on their return to homewaters.

1.2.3. The Greenland fishery

Limited fishing at West Greenland is reported as far back as the early 1900s, although the present fishery dates from 1959 when local fishers began setting fixed gillnets from small boats in certain fjords around Maniitsoq (Shearer, 1992). Rapid expansion along the coast followed and from the mid-1960s Faroese and Norwegian fishers introduced offshore driftnets, followed soon by fishermen from Greenland and Denmark. At around the same time improvements in gear (the introduction of light monofilament nets) enabled fishing in daylight and improved the efficiency of the gear. As a consequence catches rose quickly reaching a peak of almost 2700 tonnes in 1971. Fishing by non-Greenlandic vessels was phased out in 1972–1975. However, the total catch remained at around 2000 tonnes until 1976 when a TAC of 1190 tonnes was set; the fishery has been regulated since this time. Small catches of salmon are also made on the east coast of Greenland although these are sporadic and restricted by the small number of communities in this area and by drifting polar ice.

Regulatory measures have been agreed by the West Greenland Commission for most of the years since NASCO's establishment (Table 1.2.3.1). These have resulted in greatly reduced allowable catches in the West Greenland fishery, reflecting declining abundance of the contributing salmon stocks. In all but two years since 1998, the fishery has been restricted to an internal-use fishery and commercial export of salmon is not permitted. Catches in the Greenland fishery are presented in Figure 1.2.3.1.

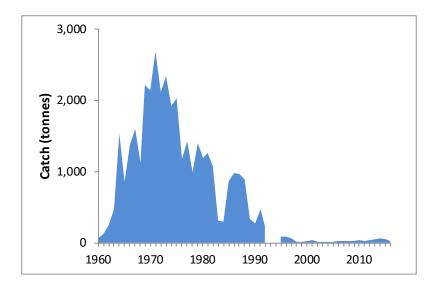


Figure 1.2.3.1. Nominal catch of salmon (tonnes, round fresh weight) in the Greenland salmon fishery, 1960–2016.

The Greenland salmon fishery operates in summer, with a fairly large proportion of the catch commonly being taken in the weeks after the opening of the season in August. Both drift and fixed gillnets continue to operate. The salmon caught in the fishery to the west of Greenland originate in both North America and the Northeast Atlantic. Data on continent of origin in the catch indicate a reasonably even split between fish from North America and Europe in the early 1990s (ICES, 2013). However, the proportion of North American fish in the catch has increased steadily since this time with North American fish comprising 80–90% of the fish caught in recent years.

The salmon caught at West Greenland are almost exclusively fish in their second summer at sea, however, these are non-maturing 1SW salmon destined to return to homewaters as 2SW, or older, fish. Fish from all parts of North America are taken in the fishery, while it is primarily only potential MSW salmon from southern countries in Europe (UK, Ireland and France) that are exploited here. Very few salmon of farmed origin appear in the catches at Greenland, and these are not taken into account in assessments.

 Table 1.2.3.1. Summary of Regulatory Measures agreed by NASCO for the West Greenland Salmon Fishery (courtesy of NASCO).

Year	Allowable catch (tonnes)	Comments/other details in the measures
1984	870	
1985	-	Greenlandic authorities unilaterally established quota of 852 t.
1986	850	Catch limit adjusted for season commencing after 1 August.
1987	850	Catch limit adjusted for season commencing after 1 August.
1988–1990	2520	Annual catch in any year not to exceed annual average (840 t) by more than 10%. Catch limit adjusted for season commencing after 1 August.
1991	-	Greenlandic authorities unilaterally established quota of 840 t.
1992	-	No TAC imposed by Greenlandic authorities but if the catch in first 14 days of the season had been higher compared to the previous year a TAC would have been imposed.
1993	213	An agreement detailing a mechanism for establishing annual quota in each of the years 1993 to 1997 was adopted by the Commission.
1994	159	
1995	77	
1996	\sim	Greenlandic authorities unilaterally established a quota of 174 t.
1997	57	An addendum to the 1993 Agreement was agreed by the Commission.
1998	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
1999	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
2000	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t. A Resolution Regarding the Fishing of Salmon at West Greenland was agreed by the Commission.
2001	28–200	Under an <i>ad hoc</i> management programme the allowable catch will be determined on the basis of cpue data obtained during the fishery.

Year	Allowable catch (tonnes)	Comments/other details in the measures
2002	20–55	Under an <i>ad hoc</i> management programme the allowable catch will be determined on the basis of cpue data obtained during the fishery.
2003–2008	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
2009–2011	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
2012–2014	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
2015-2017	Internal consumption fishery only. Greenland unilaterally committed limit the total annual catch for all components of the fishery to 45 t in 2015, 2016 and 2017.	The fishery will open no earlier than 1 Angust and close no later than 31 October each year. Any overharvest in one year will result in an equal reduction in the catch limit the following year. Efforts will be made to identify and implement temporal or spatial harvest restrictions that would provide increased protection for weaker stocks. Greenland will further improve the monitoring, management control and surveillance of its salmon fishery in accordance with the Plan for Implementation and Control Measures in the Salmon Fishery at West Greenland with the objective of achieving full catch accountability. All Members of the Commission will implement the six tenets. Greenland will inform NASCO, in a timely manner, of any modifications to the management of the West Greenland salmon fishery, of the outcome of the 2015, 2016 and 2017 fisheries and of progress with the implementation and effectiveness of its Plan for Implementation of Monitoring and Control Measures in the Salmon Fishery at West Greenland. States of origin will explore opportunities to share experiences with Greenland on monitoring, management control and surveillance in the salmon fishery. The Commission agreed to review the measure prior to the 2017 fishery.

1.3. Ecosystem aspects

Over the past 20 to 30 years, there has been a marked decline in the abundance of Atlantic salmon across the species' distributional range. Wild Atlantic salmon populations are declining across most of their home range and, in some cases, disappearing (ICES, 2008). Generally, populations on the southern edge of the distribution seem to have suffered the greatest decline (Parrish *et al.*, 1998; Jonsson and Jonsson, 2009; Vøllestad *et al.*, 2009), which may be linked to climatic factors. The decline in salmon abundance has coincided with a variety of environmental changes linked to an increase in greenhouse gases and a corresponding increase in temperatures (IPCC, 2001), which is most likely to have manifest effects at the edge of the species range. However, these areas are often also the ones with higher human population densities and therefore, typically, where potential impacts on the freshwater environment may also be greater. A range of factors in freshwater are known to affect stocks including, for example, contaminants, river obstructions, and changing river flows and temperatures (ICES, 2009b; 2010b; Russell *et al.*, 2012). Such factors have potential implications for the survival of juvenile salmon and their resulting fitness when they migrate to sea as smolts (e.g. Fairchild *et al.*, 2002).

Atlantic salmon occupy three aquatic habitats during their life cycle: freshwater, estuarine and marine. Similar factors contribute to mortality in each of these habitats competition, predation and environmental factors - but despite occurring in different habitats, these are not independent. Conditions experienced within the freshwater environment can affect the survival of emigrating smolts and marine conditions may subsequently modify the spawning success of fish in freshwater.

The decline in salmon populations has occurred despite significant reductions in exploitation, although this does not preclude possible fishery effects. An underlying cause has been a marked increase in the natural mortality of salmon at sea; the proportion of fish surviving between the smolts' seaward migration and their return to freshwater as adult fish (e.g. Peyronnet *et al.*, 2008; Chaput, 2012). For many stocks, return rates are now at the lowest levels in the time-series, even after the closure of marine fisheries. This reduced survival is thought to reflect climatic factors and broad-scale changes in ocean ecosystems as well as factors in freshwater. The exact processes controlling marine survival are relatively poorly understood (Friedland, 1998), although there is growing support for the hypotheses that survival and recruitment is mediated by growth during the post-smolt year, for European stocks at least (Friedland *et al.*, 2009).

Although their habitats are widely separated geographically, there is strong coherence in recruitment patterns between North American and European stock complexes. Recent research suggests recruitment is correlated with ocean temperature variation associated with the Atlantic Multidecadal Oscillation (AMO) (Friedland *et al.*, 2013). It further appears that there are differences in the mechanisms affecting stocks in the Northwest and Northeast Atlantic, with ocean climate variability during the first spring months of post-smolt life most important to the survival of North American stocks, while summer climate variation appears to be more important to adult recruitment variation for European stocks (Friedland *et al.*, 2013). It has been speculated that this may be related to the varying roles of predation pressure and size-related mortality on the two continental stock complexes.

In addition to changes in climate and potential issues operating in both freshwater and marine environments, various other factors have been postulated as possibly contributing to the decline in stock abundance, including predation, aquaculture impacts and the effects of fisheries. Huge increases in aquaculture production of Atlantic salmon over recent decades (see Section 2.2.1 of the WGNAS report) have created some concerns for wild populations. The main potential impacts include: (i) genetic impacts on wild fish; (ii) discharge of organic material and other wastes; (iii) transmission of diseases and parasites (particularly sea lice) to wild populations; and (iv) concerns about obtaining adequate feed resources from an already heavily exploited marine ecosystem. For example, recent investigations in Norway have demonstrated that gene pools of wild salmon populations in a number of rivers have been gradually changed through introgression of genetic material from escaped farmed salmon (Glover et *al.*, 2012; Glover et *al.*, 2013). Sea lice also continue to be regarded as a serious problem for wild salmonids (Skilbrei *et al.*, 2013; Krkošek *et al.*, 2013) affecting their survival and perhaps also their life-history characteristics (Vollset *et al.*, 2014).

As well as declines in abundance, changes in salmon life histories are also widely reported throughout the geographic range of the species, affecting factors such as sea age composition, size at age, age-at-maturity, condition, sex ratio and growth rates (e.g. Nicieza and Braña, 1993; Hutchings and Jones, 1998; Niemelä *et al.*, 2006; Peyronnet *et al.*, 2007; Aprahamian *et al.*, 2008; Todd *et al.*, 2008). Changes are also manifest in freshwater stages, affecting factors such as the size and growth of parr and the age of smolting (e.g. Davidson and Hazelwood, 2005; Jutila *et al.*, 2006) and run timing (Kennedy and Crozier, 2010; Otero *et al.*, 2013).

2. Data

2.1. Introduction

Assessment of Atlantic salmon differs from the approaches commonly adopted for other species, for example in respect of the need for at sea surveys and collection of commercial catch per unit of effort (cpue) data. Instead, the assessment of salmon is based mainly on data collected on individual river stocks (e.g. catches and counts of returning fish), which are raised and aggregated to provide estimates of the number of fish returning to homewaters for different stock groupings. These estimates are used, in turn, to estimate abundance at earlier points in the life cycle of the fish and to inform the development of catch advice.

The provision of management advice for the mixed-stock fisheries at Faroes and West Greenland is based on assessments of the status of stocks at broad geographic scales. The North American Commission (NAC) area is divided into six management units, and the Northeast Atlantic (NEAC) Commission area is divided into 19 regions. Assessment of the status of the stocks in these areas is based on estimates of the total abundance - the pre-fishery abundance (PFA) - of different cohorts of salmon at a stage before the distant water fisheries operate. PFA is defined as the cohorts of salmon maturing as 1SW and MSW fish that are alive prior to all the marine fisheries for 1SW salmon (Rago *et al.*, 1993a). The catch advice for the NEAC area is then provided for the northern (N-NEAC) and southern (S-NEAC) stock complexes and for countries.

The models to estimate the PFA of salmon from different areas are typically based on the catch in numbers of one-sea-winter (1SW) and multi-sea-winter (MSW) salmon in each country or region, which are then raised to take account of estimates of nonreported catches and exploitation rates on the two age groups. In some cases, particularly in the NAC area, returns to homewater are estimated by alternative methods, such as counts at fishways and counting fences, or from mark and recapture studies. The estimates of fish numbers returning to homewaters are then raised to take account of the natural mortality (M) between the date that the fish are deemed to recruit to the particular fishery of interest and the midpoint of the timing of the respective national fisheries. A value of 0.03 per month is assumed for M (Section 3.2.3). The date of recruitment of NAC stocks (and thus the PFA date) is taken as August 1st in the second summer at sea because these fish are first exploited in the distant water fishery at West Greenland. However, NEAC stocks recruit to the Faroes fishery during their first sea winter and so PFA is calculated at January 1st (i.e. eight months earlier) for these stocks.

2.2. Input data for assessments-NEAC area

PFA for NEAC stocks is estimated using the run-reconstruction approach described by Potter *et al.* (2004). The model estimates the PFA of both maturing and nonmaturing 1SW salmon because both stock components may be caught in the Faroes fishery, and data for both the Faroes and West Greenland fisheries are incorporated into the model.

In order to run the NEAC PFA model, each country provides time-series (beginning in 1971) of catch in numbers, non-reporting rates and exploitation rates for 1SW and MSW salmon. Best estimates and a measure of the uncertainty or error are provided for the non-reporting and exploitation rate data in order to obtain a measure of the uncertainty in the PFA estimates, since these data are commonly derived from expert opinion. The latest data input variables used in running the NEAC assessment are listed at Appendix 3.

In some instances, the above information has been supplied in two or more regional blocks per country. In these instances, the model output is provided for the regions and is also combined to provide one set of output data for the country as a whole.

The input data for Finland consists solely of catches from the River Tana/Teno. These comprise both Finnish and Norwegian net and rod catches, as the river marks the border between these countries. The Norwegian catches from the river are not included in the input data for Norway.

Where possible, when the input data are themselves derived from other data sources, the raw data are included in the model. This allows the uncertainty in these analyses to be incorporated into the modelling approach. Thus, the catch and sample data used to estimate the catches of Scottish fish in the northeast English coastal fishery are incorporated into the assessments for both UK (England and Wales) and UK (Scotland). For Greenland, catch data are input in the form of harvests (reported and unreported) in weight, along with data from the West Greenland sampling programme.

Descriptions of how the model input data have been derived are presented below for different countries (updated from Crozier *et al.*, 2003; ICES, 2002). The methods used to derive the PFA input data for NEAC countries and options for improving the data are also discussed in Crozier *et al.* (2003).

2.2.1. Median dates of return to homewater fisheries

NEAC stocks recruit to the Faroes fishery during their first sea winter and so the date of recruitment (and thus the PFA date) is calculated at January 1st. In deriving PFA from the estimates of fish numbers returning to homewaters, it is necessary to take account of natural mortality between the date that the fish recruit to the particular fishery of interest and the midpoint of the timing of the respective national fisheries. The median return date for 1SW and MSW fish for each country/region are provided in the table below. Thus there is about a six to nine month period between the PFA date and the median time of return to homewaters for maturing 1SW fish and 17 to 20 months for non-maturing fish.

NEAC Country/ region	1SW	MSW
Northern NEAC		
Russia - Pechora River	8	8
Russia - Archangel / Karelia	7.5	8
Russia - Kola / White Sea	8.5	7.5
Russia - Kola / Barents Sea	7	6.5
Finland	6.5	6
Iceland - north & east	7	6
Norway	8	5
Sweden	8.5	6.5
Southern NEAC		
Iceland - south & west	7	6
UK (Scotland - east)	8	5
UK (Scotland - west)	8	7
UK (N. Ireland - Fo area)		5.5
UK (N. Ireland - FB area)	6.5	6
Ireland	8	5
UK (England & Wales)	8	5
France	8.5	8.5

Table 2.2.1.1. Midpoint of recruitment to homewater fisheries for NEAC countries/regions.

2.2.2. Data inputs for Northern NEAC countries

2.2.2.1. Finland

Catch: The catch input to the model of Finland represents an estimate based on catch enquiries and the total number of licences issued. The Norwegian catch from the River Teno has been included in the Finnish catch, which results in a set of input data that effectively represents a single river system. Catch composition is estimated based on catch samples and corresponding scale analyses.

Level of unreported catch: Unreported catch is estimated by extrapolating the catches of the fishers that failed to report their catches, as reporting is not mandatory.

Exploitation rates: Exploitation rates in the river fisheries are derived from radio tagging studies in 1992–1993 and 1995, when 70–100 adult fish (1SW and MSW) were tagged yearly in the estuary. Most of the important river fisheries were covered by these experiments.

2.2.2.2. Norway

Area split: Salmon catches in Norway are split into four regions on the basis of climatic differences and oceanographic differences among the areas. The areas are: (1) southeast Norway from the Swedish border to the border between Rogaland and Hordaland counties, (2) southwest Norway from the border between Rogaland and Hordaland counties to Stad (3) mid Norway from Stad to Lofoten, and (4) north Norway from Lofoten to the border with Russia.

Catch: Nominal catches of salmon in the four regions are used. In recent years there have been improvements in declaring catches. From 1979 there was a weight split

1SW/MSW (<3 kg/>3 kg). From 1993 the split was changed to 1SW/2SW/3SW (<3 kg/3–7 kg/>7 kg). Mean weight was provided for most groups and used to estimate numbers in the early part of the time-series, but in recent years the reported nominal catch (reported number of killed salmon in river and sea fisheries summed) is being used. In the input to the PFA model salmon smaller than 3 kg are regarded as 1 SW fish, whereas salmon larger than 3 kg are regarded as MSW fish. The two largest size groups are thus summed into MSW salmon. In the PFA model input the Norwegian catch data for the River Teno have been removed from the Norwegian catches and incorporated in the Finnish catches.

Unreported catch: No systematic effort is used to estimate unreported catches. Inputs are guesstimates based on occasional reports from test fishing, surveillance reports, and questionnaires. Currently there is no evidence that the level of unreported catches differs between the four regions. These estimates are provided by the management authorities.

Exploitation rates: The rates for the national model are guesstimates. For parts of southeast and southwest Norway they are derived from estimated marine exploitation rates from the River Imsa and the River Drammen. In recent years (from 2009 onwards) exploitation rates for many rivers (>50) have been taken into consideration. These exploitation rates have been obtained using a multitude of methods, mainly from drift counts of spawners or results from counting facilities combined with reported catches in the rivers. The exploitation rates have been adjusted in relation to reduced fishing effort. At present different exploitation levels are used for the different regions, reflecting different harvest regimes in the regions.

For Norway, only data from 1983 onwards have been used for assessment purposes.

2.2.2.3. Russia

Area split: The Atlantic salmon rivers of northwest Russia are split into the following four regions: Kola Peninsula - Barents Sea basin; Kola Peninsula - White Sea basin; Archangelsk Region and the Karelia; and the Pechora River region. The split is based on four regions with separate catch statistics and different biological characteristics of the stocks. For example, the difference in age composition and relative abundance of summer and autumn salmon evident among these four regions has influenced the split.

Catch: The declared catch data, in numbers, is available for the full time period (1971 onwards) for all four regions. Catches were allocated to 1SW or MSW age groups on the basis of commercial and scientific catch sampling programmes.

Level of unreported catch: Unreported catches in legal fisheries are estimated from logbooks and catch statistics, by comparing catch survey results with reported catch. Illegal catch is guesstimated and based on local knowledge of fisheries. The major component of the illegal catch in the Barents Sea basin (Kola Peninsula and Pechora River) comes from in-river fisheries and a considerable part of the illegal catch in the White Sea basin (Kola Peninsula and Archangelsk region) comes from coastal areas and this contributes the greatest uncertainties. There is a particular problem with illegal catches on the Pechora River where scientific sampling programmes suggest that the illegal catch on this river is very high. The level of non-reporting increased considerably in the early 1990s due to the economic changes in Russia and temporary reduction of control and enforcement. Since late 2000s the higher level of non-reporting occurred in recreational fisheries due to unclear legislation for reporting.

All these factors have been considered in deriving the level of unreported catch for the PFA model.

Exploitation rates: Information on exploitation rates is derived from several fisheries in the Kola Peninsula where counting fences are operated and from mark–recapture exercises on the rivers with recreational fisheries. Exploitation rates in Archangelsk and Pechora are guesstimated. These are the basis of the inputs to the model, regional sea age differences being adjusted on the basis of local knowledge from estimated stock levels.

2.2.2.4. Sweden

Catch: The catch input to the model is based on annual reported commercial salmon catch on the Swedish west coast, and on voluntary reporting from sport fishing in rivers. This reporting is detailed and considered accurate and is handled by the government agency "Swedish Agency for Marine and Water Management" (commercial catches) and the Swedish University of Agricultural Sciences (non-commercial catches). Unfortunately, reporting of catches from non-commercial fishing for salmon with gillnets or rod and line on the coast is lacking. However, due to fishing regulations these catches are small (permits required for trapnets, ban on gillnets in deeper waters, restrictions on the use of gillnets in shallow waters, limited fishing season, large marine protected areas, ban on selling fish, etc.).

There is a large proportion (ca. 60–70%) of reared fish in catches and stocks as a result of compensatory releases of reared smolts (ranching). As all ranched salmon are finclipped the catches of reared fish can be treated separately in the catch statistics. In the reporting from the commercial fishing the catch is not separated into wild and reared fish. The proportion of wild salmon is instead estimated from catch statistics in nearby rivers. Stocking of reared salmon is done in three rivers; all of these also have wild stocks in tributaries.

Catch-and-release is practised in most rivers (only rod-and-line fishing allowed in rivers) but the extent of C&R is not always known. Reported catch is landed fish, but in most rivers a proportion of the fish is released back alive and any subsequent mortality is not accounted for.

Level of unreported catch: Unreported catch, i.e. non-commercial catch of salmon in the coastal area with gillnets and rod and line, is estimated from guesstimates based on expert judgement from regional fishery officers and the Swedish University of Agricultural Sciences. These estimates are supported with catch inventories carried out in 1999 (Thörnqvist, unpublished), 2004 (Swedish Agency for Marine and Water Management), 2008 (Thörnqvist, unpubl.). Generally, the unreported catch is estimated to be 5–10% of the reported catch.

Exploitation rates: Few fish counters are present and tagging data exist mainly for reared stocks, where the fishing pressure is higher than for wild stocks. Input for the PFA model is based on guesstimates. In the index River Ätran, data on size and composition of the spawning run and estimates of exploitation are developed at present. Since 2000, a fish ladder with an automatic counter has provided data on the spawning run in this river. Counter data in combination with results from small-scale tagging in this river are used to provide estimates of exploitation rates. One problem is that exploitation rates differ considerably between rivers. During the period 2000–2014 the average exploitation rates for the Swedish stock as a whole have been estimated at 34% for 1SW and 39% for MSW. The exploitation rate has increased in 2011-

2014 due to increased gillnet fishing on the coast. This fishery has been closed and exploitation is expected to decrease in future.

2.2.2.5. Iceland

Area split: The input data for the PFA model is divided into two areas. Rivers in the west and south of Iceland are combined into one area and rivers in the north and east into another. This is done on the basis of historic tag recoveries in ocean fisheries (which occurred in different areas) and different climate and oceanic conditions affecting the salmon life cycle, e.g. run-timing, smolt age, and sea age. The southern and western parts of Iceland fall within the NEAC southern area, while the northern and eastern parts of Iceland fall within the NEAC northern area.

Catch: Age-class information is available from individual catch records from logbooks used in the rod fishery. The division into sea age classes is based on a bimodal weight distribution. The 1SW females are <3.5 kg and 2SW females >3.5 kg, while 1SW males are <4 kg and 2SW >4 kg. Scale analyses have shown that the presence of salmon having spent more than two winters at sea and of previous spawners is uncommon and that the categorisation into 1SW and 2SW age classes by weight is regarded as fairly accurate. The net catches are recorded on a daily basis. The age split in the net fishery is derived from the weight distribution in the rod fishery from the same river system or from rivers in the same area.

In the River Ranga in southern Iceland substantial smolt releases have occurred since the early 1990s and have now reached a level of 300 000 to 500 000 smolts annually. Originally, the River Ranga had a small salmon stock with an annual catch of 10 to 90 fish until 1990. The river has very limited habitat for salmon production, but these 'ranched' fish now support a substantial rod fishery. The catch in the River Ranga comprised 23% (18–27%) of the total reported salmon rod catch in Iceland between 2009 and 2013. Since these fish are expected to have very low spawning success in the river they are excluded from the PFA catch input data.

Level of unreported catch: The fishing rights in Icelandic salmon rivers belong to landowners who must, by law, form a fishery association to manage the fishing right. The rod fishing rights are leased to the highest bidder. No ocean or estuary fisheries are allowed. The unreported catch was originally believed to be low with a guesstimate value of 2% applied. With increased use of midwater trawls in pelagic fisheries off the coast of Iceland, new information was provided which suggested an increased level of salmon bycatch. Based on a questionnaire survey, the value of unreported catch was therefore revised after 1995 to a value of 10% of the declared salmon catch. However, more recent analyses of DNA, as well as scale analyses, from salmon sampled as bycatch by Icelandic fishing vessels, indicates a low percentage of Icelandic salmon. Based on this, and other available information, a new estimate of unreported catch is now applied for Iceland at 4% of the declared catch for 1SW and MSW salmon.

Exploitation rates: Rates of rod exploitation are based on rivers with fish counters and catch records from logbooks. The estimates of exploitation are 40–50% for 1SW salmon and 50% to over 70% for 2SW salmon. The exploitation estimate for an inriver gillnet fishery is 39% to 52%, with a higher exploitation rate on larger fish. Information on the number of fish subject to catch and release in rod fisheries is also available from logbooks. The proportion of released fish has been increasing since 1996. The reduced exploitation due to catch and release is taken into account in the

annual estimate of exploitation for both 1SW and 2SW stock components in the PFA model inputs.

Median return date of 1SW and MSW: Run timing can vary both between years and between areas. The median return date of 1SW and 2SW salmon in south and west Iceland is mid-June and early June respectively. The median date of return is later in the north and east of Iceland, mid-June for MSW and early July for 1SW salmon.

2.2.2.6. Denmark

The Working Group collects and routinely reports the annual catch of salmon taken in Denmark. However, the small Danish catches are not included in the assessment process used in developing catch advice for the distant water fisheries.

Catch: The catch input is based on continuously collected reports of salmon taken in the recreational fishery in Danish west coast streams (from Internet sources), which all hold populations of wild salmon. In four of these, where salmon populations have always been found, there is a large proportion of reared (finclipped) salmon in the catch, but these are all F1 offspring from the original populations. In the one catchment in eastern Denmark (Gudenå), where the salmon population is not genetically native to the stream, the annual catch is guesstimated.

Level of unreported catch: Unreported catch is expected to be negligible in the western streams because the fishing is closely regulated and controlled by the anglers. In the eastern stream (Gudenå) unreported catch is guesstimated.

Exploitation rates: Exploitation rates may be derived from the total catch related to estimates of the total run (calculated by mark–recapture surveys on a three-year cycle in the four streams with original populations on the west coast).

2.2.3. Data inputs for Southern NEAC countries

2.2.3.1. France

Catch: The estimation of salmon catch in France comes from two main sources: (1) mandatory declaration of rod and line catches and from the Adour nets operating in the lower river (scales are sampled from each fish caught) to the Office National de l'Eau et des Milieux Aquatiques (ONEMA), under the Ministère de l'Ecologie, which assumed responsibility from the Conseil Supérieur de la Pêche (CSP) in 2006; and (2) mandatory declaration of catches made by professional net fishers to Affaires Maritimes, under the Ministère de la Mer, who since 2008 have delegated responsibility for collection and first processing of catch data to the Regional Boards for Sea Fisheries and Aquaculture Catch. At the same time, catches at sea are declared to the Institut Français de Recherches pour l'Exploitation de la Mer (Ifremer), who are responsible for archiving and scientific processing of all fisheries data. Salmon and sea trout catches have not been reliably collated and made available until recently. Since 1985, the 1SW/MSW split has been based on scale interpretation of the in-river catch (based on scale reading) and on a categorisation based on length thresholds for catches in estuaries and at sea. The figures prior to 1985 are not considered as reliable as the later ones.

Level of unreported catch: Unreported legal catch for the rod and line fishery has been estimated by catch inquiries made by environmental inspectors of ONEMA on each river. These procedures are still operating in some areas, but estimates are considered less reliable in recent years. The estimation of the professional net fishery catch (Adour Basin) is thought to be reliable and no unreported legal catch is considered to apply.

For most years, the unreported illegal catch is not assessed and a minimal value is provided on a precautionary basis. This unreported illegal catch has been assessed in some years by *ad hoc* inquiries in the estuary of a number of rivers in Brittany (e.g. in 2001) and on the coast (e.g. Baie of Mont Saint-Michel in 2000). The "unreported catch" is included in the nominal catch. No estimates of unreported catch are available for the early part of the time-series (prior to 2001). Thus, the rates input to the model for 1SW and MSW for the early period are near zero and range from -0.00001 to 0.00001. Higher values in the range 20% to 40% for 1SW and 15% to 30% for MSW fish are applied more recently.

Exploitation rates: Exploitation rates are derived from the index River Scorff in Brittany. This is an in-river rate, by rods only, where there are no, or very few, fish thought to be caught on the estuary or coast. Rates are also derived for the Adour river system, where a rough estimation is provided by using the lower values of adult run estimates through facilities in the three rivers flowing to Adour, and the declared catches on the coast, estuary and river, respectively by nets and rods. Some caution is necessary regarding these rates from the Adour given the uncertainties in the different estimates. The rod catch on the index River Nivelle is very small and the probable net exploitation in the estuary and coast is unknown, so exploitation rates are not used for this system. Some data on exploitation rates are also collected by ONEMA on the index River Bresle, but sea trout are the dominant angled species in this river.

2.2.3.2. Ireland

Catch: The data are derived from annual declared catches within fisheries districts, management units implemented by Regional Fisheries Boards. Since 2007 river and estuarine specific angling and commercial catch data have been complied. The Fisheries Boards were amalgamated into a single body, Inland Fisheries Ireland, in 2010 which currently takes responsibility for compiling catch statistics. Catches are split by age on the basis of a reported age distribution from 1980 to 1988. In the absence of any other information the mean proportion of 2SW salmon in the series (7.5%) has been used since 1988 and a mean of 10% has been used prior to 1980. Since the introduction of a carcass tagging and logbook scheme for angling and commercial fisheries in Ireland in 2002, sea age classes in the time-series since 2007 have been determined based upon catch dates and weights in accordance with national river stock assessments. The catch is not corrected for returns from releases of smolts for ranching or enhancement but these are not a major component of the catch.

Level of unreported catch: The values are guesstimated from local reports and knowledge achieved during catch sampling and fisheries protection activities.

Exploitation rates: A coded-wire tagging (CWT) programme has been operated in several rivers in Ireland since 1980. Up to 300 000 hatchery smolts and up to 5000 wild smolts are tagged and released annually. There is also a substantial dataset on wild salmon from the monitored River Burrishoole, providing a further index of wild returns and exploitation rates. Overall, there are estimates of exploitation rates available for three wild stocks and seven hatchery stocks for both 1SW and 2SW salmon. Up to the closure of the marine mixed-stock fishery in 2006, the annual mean of the 1SW wild exploitation index is used as the input data for the lower range of exploitation in the PFA model while the mean of the 1SW hatchery index is used as the upper range. The annual mean of the 2SW wild and hatchery exploitation index was used as

the input data for the upper and lower range of exploitation in the PFA model depending on which is higher or lower in that year. Since 2006 the main exploitation input has been from the rod catch which is estimated from coded wire tags estimates for some rivers and also rivers with counters.

2.2.3.3. UK (England & Wales)

Catch: Nominal catches for UK (England & Wales) have been derived from the catch returns submitted by netsmen and anglers and split into 1SW and MSW categories using two different methods. Since 1992, monthly age–weight keys derived from salmon caught in the River Dee trap (an index river) have been used to estimate the age of all rod-caught fish where a weight and date of capture have been provided. This has then been scaled up to the total catch (rods and nets combined) on a pro-rata basis. In earlier years (1971–1991), the age composition of the total catch has been estimated using the mean weight of the fish caught and the mean weight of 1SW and MSW salmon recovered in tagging programmes.

As the contribution of farmed and ranched salmon to the national UK (England & Wales) catch is negligible, the occurrence of such fish is ignored in the assessments of the status of national stocks. However, a large proportion of the fish taken in the northeast coast fishery are destined for Scottish rivers, and these are deducted from the returning stock estimate for UK (England & Wales) and added to the data for UK (Scotland) in the ICES assessment. This proportion is estimated to have declined from 95% of the northeast net catch in the early part of the time-series to 75% in the late 1990s and to around 65% since 2003. This reflects both the steady improvement in the status of the stocks in northeast England and the phase out of the English driftnet fishery since 2003.

Level of unreported catch: All licence holders are required to provide the Environment Agency with details of their catch of salmon and migratory trout and the number of days fished on each river or, for nets, each fishery at the end of the season. Catch returns are received from all net licence holders and from ~90% of full season anglers, and the latter account for the majority of fish caught in a catchment, typically 96–98%. The main correction for underreporting is therefore currently made in respect of perceived inaccuracies in the returns, although more substantial corrections have applied in the past.

There are few independent measures of underreporting in the rod fishery, but these indicate that the level is currently small. A value of 10% is applied for correction purposes based on the method of Small (1991). Historically, underreporting was a much more serious problem. As a result of changes in the licensing and associated catch return system covering UK (England & Wales) in the early 1990s, the percentage of underreporting in the rod catch was estimated to have decreased from ~ 50% to ~20%. Since the mid-1990s, awareness campaigns and enhanced catch reminder systems have further reduced underreporting to the levels currently estimated.

For the net fishery, a figure of 8% has been used in recent years to adjust for the level of underreporting, based on the outcome of surveillance operations. The level may have been substantially higher in the past in certain fisheries, possibly as much as 50%. However, following the successful introduction of logbooks and a carcass tagging scheme in 2009, there is now considered to be minimal underreporting in net fisheries. A figure of 2% has been assumed since 2009.

An earlier questionnaire survey of Environment Agency enforcement staff suggested illegal catches were around 12% of the declared net and rod catch. However, since the

introduction of a carcass tagging scheme and a ban on the sale of rod caught fish in 2009, it has been substantially more difficult to dispose of illegally caught fish. Since this time, illegal catches have been estimated to have been reduced to 6% of the declared catch.

Exploitation rates: Exploitation rates for a number of monitored fisheries in UK (England & Wales) are derived annually. National exploitation rates have then been estimated by deriving time-series of 'standard fishing units' employed in the salmon fisheries for the period 1971 to the present. For the period 1971 to 1997, these are calculated from the numbers of licences issued weighted by their relative catching power, which is estimated from historic cpue data; and for the period 1998 to the present, they are calculated from the numbers of days fished by different net categories weighted in the same way. The annual exploitation rates are then estimated by referencing the number of 'standard fishing units' employed over the two periods relative to average age-specific exploitation estimates derived for the 1997 and 1998 seasons.

Additional information: Further details on the derivation of estimates within UK (England & Wales) are available in the annual stock status reports (e.g. Cefas, Environment Agency and Natural Resources Wales, 2014), available at:

http://www.cefas.defra.gov.uk/publications/files/SalmonReport2013-final.pdf

2.2.3.4. UK (Northern Ireland)

Area split: Originally, a single assessment was carried out for UK (Northern Ireland). However, the data used were derived from two fishery management areas (Foyle and Fishery Conservancy Board (FCB) areas), which publish separate catch statistics and have differing fishing regulations. On the basis that stock status in the two areas may differ (Crozier *et al.*, 2003) the two areas were assessed separately from 2001.

Catch: As no commercial fishing has been conducted in the Loughs Agency area since 2010 and in the DCAL area since 2012 the Northern Ireland catch statistics currently (since 2014) rely solely on rod catches. Overall UK (Northern Ireland) rod catch estimates are available since the introduction of a carcass-tagging scheme in 2001. These catch statistics are used as an input in the model. Estimates of sea age composition of the catch for the time-series are based on 1SW/MSW data from adults returning to the River Bush (an index river).

Level of unreported catch: Estimates of unreported catch, as a result of illegal fishing, are based on intelligence reports from DCAL and Loughs Agency fishery officers. These are guesstimates only, with no verification possible. Annual adjustments in unreported catches have been used since tagging programmes started in the mid-1980s. Prior to that, a constant under reporting figure is used, as no annual data are available. The introduction of the carcass tagging scheme in 2001 has led to a reduction in unreported catches.

Exploitation rates: Estimates of exploitation rates were historically based on the River Bush microtagging programme. Exploitation from this monitored river (which is in the DCAL fishery area) was used as an input figure for all UK (Northern Ireland) fisheries (Foyle and DCAL areas). However, as currently no commercial fishery for salmon exists in the DCAL and Loughs Agency areas, exploitation rates are based on rod exploitation in the DCAL and Loughs Agency alone.

Possible improvements: A possible improvement would be to have better data available on sea age composition of all Northern Irish fish. Currently the River Bush estimate is applied to all Northern Irish data, but independent data from large river systems like the Bann and Foyle would result in a more reliable country wide estimate. In addition, a higher return rate for the carcass tagging scheme would result in more reliable estimates of exploitation rates. Recently the carcass tagging return rate in UK (Northern Ireland) has varied between 14% and 55%.

2.2.3.5. UK (Scotland)

Area split: The country is divided into eleven regions for the purposes of collating and publishing salmon and sea trout fishery statistics (Marine Scotland Science, 2012). Within the PFA run-reconstruction model, UK (Scotland) is divided into two broad areas (east and west), the split being influenced by the contrasts in climate, river size and the timing and sea age of returning fish. The east grouping comprises the east, northeast, Moray-Firth, and north statistical regions, the remaining regions comprise the west grouping.

Catch: Catches were collated according to the area split defined above. Reported retained catches of wild salmon, taken by both net and rod fisheries, are provided separately for two age classes, one sea-winter and multi sea-winter fish. Catch sampling programmes have shown that there is a variable (by region, year, and fishery) proportion of 1SW salmon categorised as MSW salmon in the reported catches. Catch data used in the model are not currently corrected to account for such misreporting.

Level of unreported catch: The ranges used in the national model are based on previous guesstimates made by local managers in some eastern areas of the country (MAFF, 1991). The differences in the ranges used for the east and west groupings are based on a subjective view that unreported catches in the west area are likely to be greater than in the east area. It is thought that human population densities are lower in the west and therefore there is likely to be less surveillance over the reporting of salmon catches. Further, west coast rivers are generally smaller and more numerous than east coast rivers, leading to a greater number of locations where unreported catches may be taken. Ranges provided are a subjective estimate of uncertainty in these parameters.

Exploitation rates. Rates for the national model are guesstimates derived from estimated exploitation rates over a range of fisheries in the early part of the time-series (Shearer, 1992), and the time-series of exploitation rates derived from the River North Esk (Marine Scotland Science, 2010), together with reported net effort indices and the proportion of rod caught fish which were subsequently released (e.g. Marine Scotland Science, 2013).

2.2.3.6. Spain

The Working Group collects and routinely reports the annual catch of salmon taken in the recreational rod fisheries in Spain (mainly Asturias). However, the small Spanish catches are not included in the assessment process used in developing catch advice for the distant water fisheries.

2.2.4. Data inputs for Faroes and West Greenland fisheries

2.2.4.1. Faroes

Reported catch: Catches are derived from the landings of salmon caught in the commercial and research fisheries that operated in the Faroes EEZ and the northern Norwegian Sea. Catches for each season (i.e. November in year n to May in year n+1) are assigned to the second year (i.e. year n+1). These fish are classified into 1SW and

MSW age groups according to their age (or potential age) on January 1st during the fishery (i.e. a post-smolt caught in November is classified as 1SW).

Unreported catch: All fish less than 63 cm total length have been discarded in this fishery and so an unreporting rate of 10-15% (with an error of +/-5%) has been used for 1SW fish; there is thought to have been negligible non-reporting of MSW fish.

Catch composition: Estimates of the proportion of farmed fish in the catch for the period 1981 to 1995 have been derived from scale reading (ICES, 1996; Hansen *et al.*, 1997); prior to 1981 all fish are assumed to have been wild, and since 1997 a value of 0.8 has been used.

Tagged fish originating in North America have been recaptured in the fishery (e.g. ICES, 1991), but excluded from catch advice.

The country of origin of the catch had been estimated based on tagging studies undertaken in the early 1990s (Hansen *et al.,* 1999). These were subsequently replaced by estimates based on genetic analysis.

Genetic investigations, based on salmon scales removed from fish caught in the fishery in the 1980s and 1990s provided an estimated proportion of North American fish in the catches at Faroes. Estimates of 5.7% (1SW) and 20.5% (MSW) have been used in subsequent analyses (ICES, 2015).

The composition of the European component was investigated using individual genetic assignments and gave an overall 1SW stock composition of 84.2% Southern European, 9.0% Northern Europe, 1.2% Icelandic and 5.7% North American (ICES, 2015). The overall composition of the MSW catch was determined as 20.9% Southern European, 58.0% Northern Europe, 0.6% Icelandic and 20.5% North American (ICES, 2015).

It was not possible to use the genetic assignments to estimate the composition of the catches to country/regional level, but they suggested that the composition within the stock complexes was broadly similar to the relative proportions of the PFA estimates and so the breakdown of catches at this level can be made by applying the relative proportions of PFA (ICES, 2015). Sources of uncertainty in these estimates are described in ICES (2015).

2.2.4.2. West Greenland

Catch: The total nominal catch (i.e. tonnes round fresh weight) in the West Greenland fishery is reported and converted to numbers using a mean weight obtained from the sampling programme.

Unreported catch: Estimates of unreported catch were not provided for the period from 1993 to 1999; an annual estimate of non-reported catch, varying from 5 to 20 tonnes was provided by the Greenland representative. Since 2000 a nominal figure of 10 t per year has been provided.

Efforts have been made to provide further information on the level of unreported catch at West Greenland. Since 2002, some assessment of the unreported catch, primarily for commercial landings, has been provided by comparing the weight of salmon seen by samplers involved in the international sampling programme and the corresponding community-specific reported landings. However, since sampling only occurs during a portion of the fishing season, these are considered to be minimum estimates for unreported catch. In addition, there is currently no quantitative approach for estimating the unreported catch for the private fishery. A telephone sur-

vey of fishers was carried out following the 2014 and 2015 seasons and provisional findings were provided to WGNAS in 2015 and 2016. These are discussed further in Section 5.1 of the Working Group report (see above); such investigations may provide a basis for revising estimates of unreported catch in future.

Continent of origin: The catch at West Greenland was divided into NAC and NEAC components using scale characteristics until around 2000 and since that time genetic analysis has been used. For the period when scale characteristics were used, the input data to the model are the minimum and maximum estimates of the proportion of NAC fish (from which minimum and maximum proportions of NEAC fish are calculated). For the subsequent period, the inputs are the numbers of NAC and NEAC fish identified in the samples.

2.2.5. Improvements to NEAC input data

NEAC countries have made ongoing efforts to improve the input data used in assessments. Modifications to input variables are reported by WGNAS in the year in which they are first implemented.

Over recent years, efforts have been made to reduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes). However, the methods used to derive estimates of unreported catch vary markedly between countries. For example, some countries include only illegally caught fish in the unreported catch, while other countries include estimates of unreported catch by legal gear as well as illegal catches in their estimates.

Descriptions of the national approaches used for evaluating unreported catches have been reported at various WGNAS meetings (e.g. ICES, 1996; 2000; 2002; 2010a). In addition, detailed reports describing national procedures for evaluating illegal and unreported catch, and efforts to minimise this, were submitted by parties to NASCO in 2007 in support of a special theme session on this issue. Full details are available at: <u>http://www.nasco.mt/pd/2007%20papers/CNL(07)26.pdf</u>.

Input data commonly rely on rod catches and the practice of catch and release has become increasingly important in recent years to reduce levels of exploitation on stocks. In the NEAC areas, catch and release estimates from the rod fisheries are not available from all countries and, when they are, corrections for catch and release mortality are commonly not applied. As the practice of catch and release is increasing, WGNAS have previously recommended (ICES, 2010a) that consideration should be given to incorporating mortality associated with this practice in river-specific, regional and national assessments.

The procedures currently used to incorporate catch and release and unreported catches into regional, national and international assessments are summarised in Appendix 1 (from ICES, 2010a).

One weakness of the NEAC model is that it is heavily dependent upon catch data and estimates of exploitation rate. In most salmon fisheries in the NEAC area, more than half the catch is reported, and in many cases it approaches 100%. However, as stocks have declined, exploitation rates have been reduced to very low levels, and estimates of abundance are therefore becoming increasingly sensitive to this parameter. This inevitably means that uncertainty in the estimates is increasing, and it therefore strengthens the need to make use of alternate sources of information on stock abundance, such as adult counts.

2.3. Input data for assessments-NAC area

The run-reconstruction model for NAC developed by Rago *et al.* (1993a) is used to estimate the PFA of non-maturing 1SW salmon of North American origin (beginning in 1971). Only the West Greenland fishery is of relevance in the context of distant water exploitation of NAC stocks. This fishery exploits predominantly (>95%) 1SW non-maturing salmon (destined to return primarily as 2SW salmon) and hence it is only necessary to estimate the abundance of this age group prior to the fishery at Greenland. The other fish taken in the fishery represent 2SW and older non-maturing salmon and previous spawners (ICES, 2003). However, PFA estimates for maturing 1SW salmon as well as large salmon (containing all MSW age groups of salmon including repeat spawners) can be derived from the run-reconstruction model.

The starting point for the reconstruction requires estimation of the returns of 2SW salmon and large salmon to the six regions in eastern North America: Labrador, Newfoundland, Québec, Gulf, Scotia-Fundy, and USA. With the progressive closure of commercial fisheries (1984 for the Gulf and Scotia-Fundy regions; 1992 for Newfoundland; and 1998–2000 for Labrador and Québec) abundance estimates have relied less on harvests and increasingly on estimated returns to rivers raised to production areas. The returns for each region are estimated with the uncertainty defined by a range of minimum and maximum values based on the best information available for each region (Chaput *et al.*, 2005).

The annual pre-fishery abundance of non-maturing 1SW fish for year i, destined to be 2SW returns (excluding 3SW and previous spawners), represents the estimated number of salmon at West Greenland prior to the start of the fishery on August 1st. Definitions of the input variables used in the model are given in Table 2.3.1. The PFA estimate is constructed by summing 2SW returns in year i+1 [NR2(i+1)], 2SW salmon catches in commercial and Aboriginal peoples' food fisheries in Canada [NC2(i+1)], and catches in year i from fisheries on non-maturing 1SW salmon in Canada [NC1(i)] and Greenland [NC1(i)].

i	Index for PFA year corresponding to the year of the fishery on 1SW salmon in Greenland and Canada
М	Natural mortality rate (0.03 per month)
t1	Time between the midpoint of the Canadian fishery and return to river = 1 month
S1	Survival of 1SW salmon between the homewater fishery and return to river $\{exp\text{-}M * t1\}$
H_s(i)	Number of "Small" salmon caught in Canada in year i; fish <2.7 kg
H_l(i)	Number of "Large" salmon caught in Canada in year i; fish >=2.7 kg
AH_s	Aboriginal and resident food harvests of small salmon in northern Labrador
AH_l	Aboriginal and resident food harvest of large salmon in northern Labrador
f_imm	Fraction of 1SW salmon that are immature, i.e. non-maturing: range = 0.1 to 0.2
af_imm	Fraction of 1SW salmon that are immature in native and resident food fisheries in N Lab
q	Fraction of 1SW salmon present in the large size market category; range = 0.1 to 0.3
MC1(i)	Harvest of maturing 1SW salmon in Newfoundland and Labrador in year i
i+1	Year of fishery on 2SW salmon in Canada
MR1(i)	Return estimates of maturing 1SW salmon in Atlantic Canada in year i
NN1(i)	Pre-fishery abundance (PFA) of non-maturing 1SW + maturing 2SW salmon in year i
NR(i)	Return estimates of non-maturing + maturing 2SW salmon in year i
NR2(i+1)	Return estimates of maturing 2SW salmon in Canada
NC1(i)	Harvest of non-maturing 1SW salmon in Nfld + Labrador in year i
NC2(i+1)	Harvest of maturing 2SW salmon in Canada
NG(i)	Catch of 1SW North American origin salmon at Greenland
T2	Time between the start of the fishery at West Greenland (August 1) and return to the coast of North America = 10 months
S2	Survival of 2SW salmon between August 1 (at West Greenland) and return to the coast of North America {exp-M * t2}
MN1(i)	Pre-fishery abundance of maturing 1SW salmon in year i

 Table 2.3.1. Definitions of key variables used in continental run-reconstruction models for North

 American salmon.

2.3.1. Data inputs for NAC

The latest data input variables used in running the NAC assessment are listed at Appendix 4. More detailed descriptions of how the model input data have been derived for each region of North America are presented below.

2.3.1.1. Labrador

For Labrador stocks, it was thought to be inappropriate to develop total recruits from angling catches and exploitation rates similar to techniques used for rivers in insular Newfoundland. The problem with using angling catches to derive returns for Labrador is, that until 1994, there were no estimates of exploitation rates available other than for the salmon population of Sand Hill River and these were 20 years out of date. Also, because Labrador coastal rivers are isolated, the exploitation rates are low and highly variable depending on the presence of an angling camp and its success in attracting guests as well as the nearness of local communities. Thus, exploitation rates would depend, and vary from one year to the next, on the success of angling camps in attracting anglers and may not be applicable to other Labrador rivers. Thus, all estimates of returns and spawners until 1998 were based on commercial catches as the only source available of usable continuous time-series of data.

Before 1998

The general approach is to use exploitation rates to convert commercial catches of small and large salmon in Labrador to total population prior to the commercial fishery. River returns and spawners were estimated by subtracting the commercial catch from these populations, and accounting for non-Labrador interceptions. The estimated number of Labrador origin large returns is calculated as:

 $LR = (CC^*PL) / u$

where,

LR = Labrador returns, PL = proportion Labrador origin, CC = commercial catch, and u = exploitation rate

The estimated number of Labrador origin small returns is determined from equation (1) but using commercial catches of small salmon.

Parameter values for sea age and the proportion of salmon of Labrador origin comes from the sampling program in the commercial fishery, 1974–1991. In 1997, commercial sampling resumed with samples being collected throughout the fishery at Makkovik and Rigolet in SFA 1 and Cartwright and St Lewis/Fox Harbour in SFA 2. River age distribution of commercial samples of small and large salmon from Labrador have been found to consist, on average, of about 75-80% river age 4 and older in SFAs 1 & 2. The commercial samples came from commercial catches sampled in Labrador at several sites along the Labrador coast including Square Islands (SFA 2) and at Nain (SFA 1) (Anon, 1993b). In total, 46 320 salmon were sampled for scales and aged. Labrador salmon stocks are thought to contribute about 70% of the total production of four year, and older, river age salmon, with the other 30% coming from northern Québec. Thus, when non-Labrador salmon are factored in at 30% applied to the river age distribution, then 60–80% of the harvest of small and large salmon (PL) in Labrador are of Labrador origin (Anon, 1993b). In 1997, in SFA 1, the percentage of the commercial catch that was of Labrador origin was for large salmon 68% (95% C. I. 64.3-72.5%); whereas for small salmon it was 39% (95% C.I. 35.6%-41.6%). In 1997, in SFA 2, the percentage of the commercial catch that was Labrador origin was for large salmon 92% (95% C. I. 88.4-95.2%); whereas for small salmon it was 80% (95% C.I. 74.8%-85.0%).

Exploitation rates (u) were calculated from the smolt tagging study in 1969–1973 on Sand Hill River (Reddin, 1981; Reddin and Dempson, 1989). Exploitation rates of 0.28 to 0.51 for small salmon and 0.83 to 0.97 for large salmon from the tagging study were changed to base exploitation rates of 0.3 to 0.5 on small salmon and 0.7 to 0.9 on large salmon and were assumed to apply to all of the salmon populations in SFAs 1, 2, and 14B for the period of 1969–1991 (Anon., 1993b). After 1991, due to the Management Plans for the commercial fishery in Labrador and Newfoundland, several changes occurred that would reduce exploitation of Labrador origin salmon. These changes include: (1) reductions in effort as commercial salmon fishers chose to sell their li-

(1)

cences from a buy-out agreement begun in 1992, (2) a moratorium on commercial fishing in Newfoundland would increase the number of Labrador salmon in Labrador coastal waters, and (3) season reductions due to the varying opening dates and early closures from the quotas applied in 1995 and 1996. The effects of these changes were quantified in the exploitation model as follows:

 $u=1-e^{-aF}$ (2)

where: a = fraction of the 1991 licensed effort remaining in 1992–1996.

In 1994–1996, the licensed effort for all of Labrador was 37% of the 1991 level of 570 licences, in 1993 it was 55%, and in 1992 it was 87%. In any given year, it was assumed that 90% of licensed fishers were active. Fishermen reported during public consultations that in 1995 and 1996 many licensed salmon fishers did not fish for salmon in 1995–1996 but fished for crab instead. This was verified by Fisheries Officers who reported that of the 218 licensed salmon fishers only 132 were active in 1996. Another method of obtaining actual effort information is also available since, beginning in 1993 commercial fishing vessel (CFV) numbers have been recorded on sales receipts issued to fishers by fish plants. Enumeration of licensed salmon fishers actively fishing was made by determining the number of CFVs in the Statistics Branch catch records. Active effort in 1991 and 1992 was assumed to be 90% as it was in 1993 and 1994 from the CFV file. Thus, the exploitation rates (u) were modified due to effort reductions in equation (2) using estimated active licences from 1991 as a base and the number of active licences in 1995, 1996 and 1997. The modified exploitation rates (ue) for 1992–1997 used the licensed effort in equation (2).

The tagging study on Sand Hill River, 1969–1973 showed that Labrador small and large salmon were not only caught in Labrador, but also in the commercial fisheries along the northeast coast of Newfoundland (both small and large) and at West Greenland (large only) (Anderson, 1985). For small salmon, out of a total of 100 (1SW) tag returns there were 24 from Newfoundland. For large salmon, out of a total of 137 (2SW) tag returns there were 41 from Newfoundland.

For 1992–1997: the moratorium on commercial fishing in Newfoundland would have released small and large salmon to Labrador. The effect of salmon released from Newfoundland in 1992–1996 was evaluated against the exploitation rates as follows:

$$un = (1-((24 * (1 - ue))/100)) * ue, for small salmon, andun = (1-((41 * (1 - ue))/137)) * ue, for large salmon (3)$$

The new estimates of fishing mortality (un) in 1992–1994 included adjustments for the closure of the commercial fishery in Newfoundland based on the results of the Sand Hill River tagging study. Season reductions due to the varying opening dates and early closures from the quotas applied in 1995 and 1996. In 1995, adjustments were made to account for the new opening date for the commercial fishery in Labrador of July 3 changed from June 20 the previous year. For 1995, the accumulative effect of these, weighted to SFA catches, was to reduce the catch so that for small salmon the current catch represents 86.0% of small salmon and 62.7% of large salmon. In 1996, the opening date reverted to June 20 but the quota levels resulted in early closures in SFA 2 of 2A - July 10, 2B - July 8, and 2C - July 2 while SFA 1 and 14B did not close. For 1996, the accumulative effect of these weighted to SFA catches was to reduce the SFA catches was to reduce the catch so that for small salmon and 61% of large salmon. In 1997, the opening date remained at June 20 but the quota levels resulted in early closures the catch so that for small salmon and 61% of large salmon. In 1997, the opening date remained at June 20 but the quota levels resulted in early closures in SFA 2 of 2A - July 12, 2B - July 15, and 2C - July 13

while SFA 1 closed on October 15 as the quota was not caught. For 1997, the accumulative effect of these early closures was to reduce the catch so that for small salmon the current catch represents 47% of small salmon and 64% of large salmon. The season changes reduce catches and hence lower exploitation rates. The effect of shorter seasons in 1995, 1996 and 1997 was evaluated against the exploitation rates in section B as follows:

The new estimates of fishing mortality including effort reductions, adjustments for the closure of the commercial fishery in Newfoundland, and shorter seasons due to opening dates and quotas results in the following exploitation rates which were applied to catches. The cumulative effect of factors A, B, and C is to reduce exploitation on Labrador origin salmon.

Labrador origin 2SW returns (LR2SW) were derived from eq. 1 b

LR2SW = LR * P2SW

where: P2SW = proportion of the large salmon that is 2SW salmon.

The SR1SW were calculated as in equation (5) but using P1SW which is the proportion of the catch that is 1-sea winter in age and maturing to enter freshwater and spawning in the year of capture. The parameter values for P1SW of 0.1 to 0.2 come from Anon. (1991).

The 2SW component was estimated separately for salmon caught in SFA 1, 2 and 14B. In SFA 1, commercial sampling at Nain of large salmon showed the proportion of 2SW were on average about 84% (n=6542), 1977–1991. Thus, a range of 0.7–0.9 was used for SFA 1. In SFA 2, commercial sampling of large salmon averaged 69% (n=4793) 2SW salmon, 1977–1991. There were no commercial samples available for SFA 14B. Thus, for SFAs 2 & 14B a range of 0.6–0.8 was used. For the 1SW component, commercial samples at Nain in SFA 1 of small salmon showed the proportion of 1SW salmon were on average about 94% (n=4757). In SFA 2 the 1SW component was on average about 97% (n=8872) of small salmon. There were no samples from commercial sampling in SFA 14B. In 1997, aged commercial samples indicated that the previous range was acceptable.

Total river returns of 2SW salmon (TRR) were calculated as follows:

$$TRR = LR2SW / (1-us)$$
(6)

The total river returns of small salmon are also calculated by equation 6 but from SR.

Spawning escapement (SE) or spawners was calculated according to the formula:

$$SE = TRR - AC, \tag{7}$$

where:

AC = angling catch which includes retained catch plus 10% of hook & released mortality for released salmon.

A couple of modifications were made to the estimation procedure for Labrador in 1997. First, determination of exploitation rates were calculated separately for SFA 1, 2

(5)

and 14B using the active effort individually for each SFA. For SFA 1, the active number of licences declined from 141 in 1991 to 39 in 1997. For SFA 2, the active number of licences declined from 320 in 1991 to 99 in 1997. For SFA 14B, active licences declined from 52 in 1991 to 0 in 1997 when the fishery was closed. Exploitation rates determined as in equations 2, 3 and 4 are: SFA 1 - small was 0.0735 to 0.1399 and - large was 0.2221 to 0.3959; and SFA 2 - small was 0.0384 to 0.0728 and - large was 0.1589 to 0.2799.

Numbers of small and large salmon for SFAs 1 & 2 were estimated from the exploitation model while for SFA 14B the results of assessments on Forteau Brook and Pinware River were expanded to include all the watersheds in SFA 14B. Returns to SFA 14B were 663 to 1545 small salmon and 146 to 327 large salmon.

Total mortalities of small and large salmon were accounted for by summing commercial catches of small salmon in Labrador and Newfoundland, large salmon in Labrador, Newfoundland, and Greenland, angling catches in Labrador of small and large salmon including 10% of the hook and released salmon, and small and large spawners. All of the above mortality estimates except catches of Labrador salmon in Newfoundland, 1969–1991 and Greenland could be obtained from equations 1 to 7. Catches in Newfoundland and Greenland were assessed as follows:

Greenland: for 1969–1992 and 1995–2004, removals of Labrador salmon by the Greenland fishery were assessed from data based on the sampling program in commercial fish plants at West Greenland (Anon, 1996). The Greenland fishery catches salmon that would have returned to homewaters as large salmon in the year following the Greenland fishery. Numbers of Labrador salmon were determined by converting catches in kg to numbers of salmon of 1SW North American origin that were of river age 4 and older. The number of Labrador salmon were estimated by assuming that 70% of the production of 4-year and older river age salmon are from Labrador (Anon, 1993b).

Newfoundland: for 1969–1991, catches of Labrador small and large salmon in Newfoundland were included in total mortalities as the product of the ratio of tags caught in Newfoundland to Labrador and the catch in Labrador. For small salmon the ratio was (24/(100-24)) = 0.32 and for large salmon it was (41/(137-41) = 0.43).

1998-2001

For the years, 1998–2001 when only one or two counting projects took place in Labrador, the raising factors previously used and explained of 1.04 to 1.49 for small salmon and 1.05 to 1.27 for large salmon were used to estimate returns and spawners for Labrador from the overall PFA minus catches in Greenland, as was the case in previous years. However, in this case returns to rivers were derived for Labrador by subtracting landings in food fisheries. Also, catches in 1994–2006 were updated to reflect changes made to catch statistics in Labrador from the Licence Stub Return System. Procedures for the collection and compilation of commercial and angling fishery data are described in Ash and O'Connell (1987) for fishery years 1974–1996. For years 1969–1974, commercial catch data came from Anon. (1978). In 1997, the angling catch statistics were converted to a licence stub system (O'Connell *et al.*, 1998) which continues to present.

2002-present

Counting projects occur on three to four Labrador rivers; out of about 100 extant salmon rivers. Because they occur on the same rivers each year, it is possible to ex-

trapolate from abundance for small and large salmon per accessible drainage areas in these monitored rivers to unsurveyed ones in the remainder of Labrador. The accessible drainage areas were 9267 km² for Lake Melville (SFA 1A), 25 485 km² for Northern Labrador (SFA 1B), 28 160 km² for Southern Labrador (SFA 2), and 2651 km² for the Straits Area (SFA 14B). Accessible drainage area in the counting facility rivers was 1878 km² resulting in an expansion factor of 35 to one. Not all rivers in Lake Melville were included due to a lack of information on presence of salmon populations in rivers in this region of Labrador. Lake Melville rivers whose drainage areas were included are Sebaskachu, Cape Caribou, Goose, MacKenzie, Kenamu, Caroline and Traverspine.

Abundances for SFAs 1A and 1B were derived from English River returns with maximum and minimum values developed using the observed variability of relative abundances in SFA 2. Total returns and spawners for Labrador are estimated by Monte Carlo simulation based on 10 000 random draws from the range of values assuming abundances per km² of accessible drainage were uniformly distributed. The relative abundances (per km²) for each SFA were then multiplied by the total accessible drainage area to derive total returns of small and large salmon. Ranges of values were developed to convert numbers of small and large salmon to numbers of 1SW and 2SW salmon from scale age information collected from counting fences and angling fisheries in Labrador. A bootstrap procedure was used to develop estimates of the proportions of sea age 1 salmon in estimates of small salmon returns and spawners, proportions of sea age 1 salmon in the estimates of large salmon returns.

Sea age correction factors were:

Small to 1SW - 96 to 100% Large to 2SW - 60 to 71% Small overlap in large - 12 to 21%

Spawners of 1SW and 2SW salmon were derived by subtraction of angling catches including an estimate of hook and release mortalities (10%) from the returns.

2.3.1.2. Newfoundland

Inputs for the run-reconstruction model for Newfoundland include estimates of small, large and 2SW returns and spawners to rivers (minimum and maximum). The methods used to estimate returns and spawners to the rivers in Newfoundland are described by Reddin and Veinott (2010). In brief, returns and spawner estimates were derived from recreational fishery exploitation rates of retained small salmon for rivers with enumeration facilities; and ratios of large to small salmon were utilized to estimate large salmon. Exploitation rates were then applied to all rivers with reported angling catches. A non-parametric bootstrap technique was used, whereby exploitation rates and ratios of large to small salmon from rivers with enumeration facilities were chosen at random with replacement. The 95th confidence interval from 500 iterations of the weighted exploitation rate and ratio of large to small salmon was applied to angling catches on a Salmon Fishing Area (SFA) basis. The midpoint of the 95th confidence interval was used as the minimum and maximum estimate returns of large and small salmon in each SFA. Estimates of 2SW returns are based on the expected proportion of 2SW in the large salmon category (≥ 63 cm). Commercial and recreational angling catches were derived as described for Labrador (2.3.1.1). Spawners in all years were determined as the returns to rivers minus angling catches including an adjustment for hook-and-release mortality.

2.3.1.3. Québec

In order to estimate abundance of stocks, rivers were classified into six categories (C1–C6) depending on the information available to estimate salmon returns (according to the method of Caron and Fontaine, 1999), with C1 being the most reliable evaluation and C6 the least. C1 corresponds to a river where the evaluation of the returns is based on a counting method, either from a fence or from a visual count through snorkelling or from a canoe. C2 uses the same evaluation, but without knowing the number of small and large salmon, which is then estimated from proportions reported in the sport fishing landings and, if necessary, the catch and release. Salmon returns on C3 rivers are determined based on multiple correlation factors, using catch number, fishing effort, season duration and river accessibility distance (Guillouët, 1993).

When estimation of the returns using a C1–C3 category is not possible, and when data of returns from previous years are available, the C4 category is used. C4 assumes that interannual variations in salmon returns in the targeted river are approximately the same as variations observed in the other rivers of the corresponding region. Category C5 is for rivers where only landings data are available. In these rivers the salmon run is estimated from the average regional exploitation rate. Finally, a few small rivers have essentially no available data. C6 then assumes that the run is related to the available river salmon habitat and is estimated with respect to rivers of the same area for which run estimates and salmon habitat area are known. Estimated numbers of returns from C4 to C6 cannot be used to assess relative to attainment of conservation limits. However, they provide at least approximate numbers to estimate returns and spawners for salmon rivers in Québec.

The evaluation of the uncertainty associated with return estimates depends on the river category. For C1 and C2 rivers, the correction factor for the minimum and maximum number of returns is +5% and +10% for all rivers with a fish ladder and for all others in zones Q1 to Q3 and Q10. The correction factor for rivers with darker water from zones Q5, Q6 and Q7 is rather +10% and +30%. For the other categories, an uncertainty of ±25% is associated with salmon return estimates, except for category C3 where calculation depends on the method of Guillouët (1993).

The number of spawners is obtained using the return estimate minus all river catches, which include landings and other types of removal. In most cases, river catches include landings from sport fishing only, which may be conducted by native people such as that on the Betsiamites River. The other types of removal are of limited number and include mainly natural mortality, salmon captured for hatchery use and subsistence fishing when practised in river.

Overall return estimates for all Québec rivers are obtained by adding in-river salmon returns, commercial fishing (when operated), native people subsistence fishing when practised in estuaries and an estimate of non-registered landings. However, little scientific data are available on non-registered landings and thus, estimates are based on good judgment, following consultations with regional biologists.

2.3.1.4. Gulf

Estimation of returns and spawners are developed for the four salmon fishing areas of Gulf Region (SFAs 15 to 18).

SFA 15

The major river in this area is the Restigouche River. The returns and spawners are estimated for the Restigouche River exclusive of returns to the Matapedia River, which are included in Québec zone Q1. The Restigouche River stock assessment is based on angling catch with assumed exploitation rates between 30% (min.) and 50% (max) with estuary catches added back after the estimates of returns. Return and spawner estimates for SFA 15 are based on Restigouche River data, scaled up for SFA 15 using angling data. The return and spawner estimates for SFA 15 are derived from the return and spawner estimates for Restigouche (New Brunswick). The minimum and maximum return and spawner estimates are derived from the minimum and maximum ratios of angling catch in all of SFA15 relative to angling catch in Restigouche (New Brunswick) (min = 1.117; max = 1.465). Harvests represent retained angling catch plus 6% catch and release mortality for released fish. The proportion of 2SW in large salmon numbers is based on aged scale samples from angling, trapnets, and broodstock. In the years when no scale samples analysis is available, a mean value of 0.65 is used.

SFA 16

The most important Atlantic salmon river in SFA 16 is the Miramichi River. The Miramichi makes up 91% of total rearing area of SFA 16 and returns to the river are assessed annually. For 1971 to 1991, minimum and maximum values are based on capture efficiencies of the Millbank estuary trapnet representing a lower CI of -20% of the estimate and upper CI of 33% of the estimate. For 1992 to 1997, minimum and maximum are lower and upper CI and based on estimate bounds of -18.5% to +18.5%. Since 1998 to the present, minimum and maximum are 5th and 95th percentile range from a Bayesian hierarchical model used in the assessment. Returns to SFA 16 are Miramichi returns (Minimum, Maximum) / 0.91. Proportion 1SW in small salmon is from scale ageing; proportions have varied from 0.97 to 1.0. Proportion 2SW in the large salmon category is obtained from scale ageing. Spawners are returns minus harvests. For 1998 to 2011, the harvest of large salmon is estimated as the sum of the aboriginal fisheries harvests for large salmon and 1% of the large salmon catch (30% exploitation rate, 3% catch and release mortality). The harvest of small salmon is estimated as 30% of the small salmon return plus the harvest from the aboriginal fisheries.

SFA 17

For 1970–1994, small returns are estimated from retained small salmon catch in the Morell River divided by the river-specific exploitation rate. Salmon catch in the Morell River was estimated in 1970–1990 by DFO Fisheries Officers; and in 1991, 1992, and 1994 by angler mail-out surveys. The number of small retained salmon in 1993 was not recorded, so the number used is the mean for 1986–1992. For 1970–1993, exploitation rate was taken as the mean of exploitation rates estimated for 1994, 1995, and 1996 (0.317). For 1994, exploitation rate was 0.34. The min and max of small returns are calculated using exploitation +/- 0.1; e.g. 0.34 +/- 0.1 gives 0.24 and 0.44. Large returns = (number of small returns/proportion small) - number of small and large salmon in the angling catch of each year. For 1981–1994, proportion small is taken from counts at the Leards Pond trap on the Morell River. Small spawners = number of small recruits-number of small retained. Large spawners = number of large recruits - number of large retained.

Spawner estimates for 1995 to the present are derived from redd counts in 23 rivers. For years and rivers in which redd counts are unavailable, redd numbers are estimated by linear interpolation from the preceding and succeeding count year. Redd numbers in years prior to the first count are taken as the first count. Redd numbers in years after the last count are taken as the last count. Female spawners are estimated from the ratio of 3.357 redds/female spawner, measured in the West River in 1990. Total spawners are estimated from size-specific sex ratios derived from counts at Leards and Mooneys Ponds, Morell River, in 1986–2001. The proportion of salmon that are large is taken as 0.5 in the Cains, Carruthers, Trout (Coleman), Morell, Cardigan, West, and Dunk Rivers, and 0.9 in all other rivers. Spawners are presented as Min (estimated spawners -20%) and Max (estimated spawners + 20%). Returns are spawners + total estimated fishing mortality, including angler catches, hook-andrelease mortality, and native harvests. Angler catches and hook-and-release mortality are estimated from angler card surveys. Returns are presented as Min (estimated returns -20%) and Max (estimated returns + 20%). It is assumed that large salmon and 2SW salmon are equivalent.

SFA 18

Returns and spawners to SFA 18 are derived from estimates of returns and spawners to the Margaree River, adjusted for the ratio of the SFA 18 angling catch to the Margaree River catch. For small salmon, the ration of SFA 18 catch to Margaree catch varies between 1.15 and 2.71 for years 1984 to 2004. For large salmon, the ratio of SFA 18 catch to Margaree catch varies between 1.08 and 2.32 for years 1984 to 2004. Returns to Margaree River are estimated using various techniques.

- 1970 to 1983 angling catch divided by range of exploitation rates with maximum exploitation rate of 0.37 and minimum exploitation rate of 0.215;
- 1984 to 1986 based on annual assessments;
- 1987 to present angling catch and effort data from logbooks and provincial licence stubs are used to derive the returns. The catchability coefficient per rod day is estimated from angling catch and effort data for the years 1988 to 1996 when mark and recapture programmes were used to estimate returns, independently from angling data.

Spawners for 1970–1983 equal returns minus removals. Spawners for 1984 to the present equal returns minus catch for small salmon and returns minus catch, corrected for 5% mortality, for large salmon. 2SW salmon represent between 0.77 and 0.87 of large salmon returns and spawners.

2.3.1.5. Scotia-Fundy

Salmon originating in rivers of the Atlantic coast of Nova Scotia and southwest New Brunswick in Salmon Fishing Areas (SFAs) 19–21 and the portion of SFA 23 outside the inner Bay of Fundy comprise the Scotia-Fundy stocks. With the exception of at least one stock in SFA 19, they have a large salmon component that migrates to the North Atlantic/Labrador Sea (Amiro *et al.*, 2008). Estimates of returns and spawning escapement for the Scotia-Fundy stocks are provided as inputs to the run-reconstruction model. Inner Bay of Fundy Atlantic Salmon (SFA 22 and part of SFA 23) have been federally listed as endangered under the Canadian Species at Risk Act and are not included as inputs into the run-reconstruction model. With the exception of one population, inner Bay of Fundy stocks have a localized migration strategy while at sea and an incidence of maturity after one winter at sea.

Consistent with the requirements of the model, a range (minimum to maximum) of returns and spawning escapement for the Scotia-Fundy stocks is provided for the run-reconstruction model. The methods used to estimate total returns and spawners are described by Amiro et al. (2008). In brief, for SFAs 19–21, the escapement is based on the count of small and large salmon at the Morgan Falls fish-way on the LaHave River from 1970 to the present year, scaled up to the region using the relationship between this count and the recreational catch data for rivers in SFA 19 to 21 from 1970 to 1997 and a catch rate for the LaHave River from 1970 to 1997. Estimates of the returns also include estimates of landings in the commercial salmon fisheries in SFA 19–21 from 1970 to 1983. The model is fitted using maximum likelihood, and the 90% confidence limits are carried forward as the minimum and maximum values. In SFA 23 from 1970 until 1992, estimates of total 1SW and large wild-origin salmon returns are based on the estimated number of returns destined for tributaries above Mactaquac Dam on the Saint John River; this includes in-river and outer-Fundy commercial landings (1970–1971 and 1981–1983), in-river aboriginal harvests (since 1974), and counts at Mactaquac Dam. These estimates are raised by the proportion of the total accessible productive habitat in SFA 23 that is upstream of Mactaquac Dam (0.4–0.6). Hatchery-origin returns were attributed to above Mactaquac Dam only and no hatchery 1SW and MSW returns were estimated for other rivers within SFA 23 (outer Fundy). Since 1993 the estimates of 1SW and MSW returns to the Nashwaak River have been used to estimate the wild production from tributaries of the Saint John River below Mactaquac Dam. The estimated 1SW and MSW returns to the Nashwaak River (above Counting Fence), is raised by the proportion of the total production area accounted for below Mactaquac (0.21-0.3) and then added to the above Mactaquac totals.

2.3.1.6. USA

Total salmon returns and spawners for USA rivers are based on trap and weir catches and for the small rivers in Maine that do not have fish counting facilities, estimates of spawners were based on redd counts.

2.3.2. Improvements to NAC input data

Modifications to input variables used in assessments for the NAC area are reported by WGNAS in the year in which they are first implemented.

Over recent years, efforts have been made to reduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes). However, the methods used to derive estimates of unreported catch vary markedly between countries. For example, some countries include only illegally caught fish in the unreported catch, while other countries include estimates of unreported catch by legal gear as well as illegal catches in their estimates.

Descriptions of the national approaches used for evaluating unreported catches have been reported at various WGNAS meetings (e.g. ICES, 1996; 2000; 2002; 2010a). In addition, detailed reports describing national procedures for evaluating illegal and unreported catch, and efforts to minimise this, were submitted by parties to NASCO in 2007 in support of a special theme session on this issue. Full details are available at: <u>http://www.nasco.int/pdf/2007%20papers/CNL(07)26.pdf</u>

Input data commonly rely on rod catches and the practice of catch-and-release has become increasingly important in recent years to reduce levels of exploitation on stocks. As the practice is increasing, WGNAS have previously recommended (ICES, 2010a) that consideration should be given to incorporating mortality associated with this practice in river-specific, regional and national assessments.

The procedures currently used to incorporate catch-and-release and unreported catches into regional, national and international assessments are summarised at Appendix 1 (from ICES, 2010a).

2.4. Biological and other data requirements

As noted previously, many of the 'conventional' data requirements (e.g. marine survey data and commercial cpue) used in the assessment of other commercially important fish species are inappropriate to salmon. A range of biological, catch and exploitation rates and other data pertinent to appropriate stock assessments are however, collected and made available to WGNAS to help inform assessments and to aid in responding to the various questions posed by NASCO.

Appendix 2 of this Stock Annex provides an overview of current and possible future data requirements for Atlantic salmon assessment/ scientific advice. This was compiled at a recent meeting of WGNAS (ICES, 2013) in relation to monitoring requirements under the European Data Collection Framework (DCF) and following a more detailed review of the data requirements under DCF (ICES, 2012c). This table illustrates the type of information collected/available, but is provided for illustrative purposes only. It should be noted that many Atlantic salmon producing countries fall outside the DCF provisions, which only relate to countries within the European Union. Further, Sovereign states are responsible for the regulation of salmon fisheries within their areas of jurisdiction. Formal ICES catch advice is only required for the distant water salmon fisheries, which take salmon originating in rivers of another party.

3. Assessment methods

In managing Atlantic salmon fisheries, NASCO has adopted a fixed escapement strategy (Potter, 2001), in recognition of the importance of the spawning stock to subsequent recruitment. Therefore, in managing the distant water fisheries at Faroes and West Greenland, the spawning requirements of the rivers contributing to these fisheries must be defined. Management advice, expressed as allowable harvest (tonnes), is then predicated on a forecast of salmon abundance prior to the fishery such that the spawning requirements of the contributing stocks can be achieved. The provision of catch advice thus proceeds through a number of steps:

The definition of spawning objectives;

The development of a measure of abundance prior to the fishery; i.e. the prefishery abundance or PFA;

A measure of the spawning stock contributing to the PFA;

A model to forecast the PFA;

The development of a risk analysis framework for the catch advice.

These steps are described in detail in the following sections, subdivided as necessary for the different distant water fisheries and the various stock complexes which contribute to the two fisheries (Greenland and Faroes).

3.1. Definition of spawning objectives

3.1.1. Management objectives and reference points

Conservation limits (CLs) for North Atlantic salmon have been defined by ICES as the stock level that will achieve long-term average maximum sustainable yield (MSY). NASCO has adopted the following definition of CLs (NASCO, 1998): 'The CL is a limit reference point; having populations fall below these limits should be avoided with high probability.'

Atlantic salmon has characteristics of short-lived fish stocks; mature abundance is sensitive to annual recruitment because there are only a few age groups in the adult spawning stock. Incoming recruitment is often the main component of the fishable stock. For such fish stocks, the ICES MSY approach is aimed at achieving a target escapement (MSY Bescapement, the amount of biomass left to spawn). No catch should be allowed unless there is a high probabaility that this escapement can be achieved. The escapement level should be set so there is a low risk of future recruitment being impaired, similar to the basis for estimating B_{pa} in the precautionary approach. In short-lived stocks, where most of the annual surplus production is from recruitment (not growth), MSY Bescapement and B_{pa} might be expected to be similar.

It should be noted that this is equivalent to the ICES precautionary target reference points (S_{pa}). Therefore, stocks are regarded by ICES as being at full reproductive capacity only if they are above the precautionary target reference point. This approach parallels the use of precautionary reference points used for the provision of catch advice for other fish stocks in the ICES area.

For the assessment of the status of stocks and advice on management of national components and geographical groupings of the stock complexes, where there are no specific management objectives:

- ICES requires that the lower bound of the 90% confidence interval of the current estimate of spawners is above the CL for the stock to be considered at full reproductive capacity (equivalent to a probability of at least 95% of meeting the CL).
- When the lower bound of the confidence limit is below the CL, but the midpoint is above, then ICES considers the stock to be at risk of suffering reduced reproductive capacity.
- Finally, when the midpoint is below the CL, ICES considers the stock to be suffering reduced reproductive capacity.

Ideally, Atlantic salmon should be assessed and managed on the basis of riverspecific stock units, the scale corresponding best to the spawner to recruitment dynamic (Chaput, 2012). In reality, this is not the case for the majority of rivers, although efforts are continuing to develop river-specific CLs and assessment protocols and developments are reported annually to WGNAS (e.g. ICES, 2013).

The risk assessment frameworks applied by WGNAS directly evaluate the risk of meeting or exceeding the stock complex objectives. Managers can choose the risk level which they consider appropriate. ICES considers however that to be consistent with the MSY and the precautionary approach, and given that the CLs are considered to be limit reference points and to be avoided with a high probability, then managers should choose a risk level that results in a low chance of failing to meet the CLs. ICES

recommends that the probability of meeting or exceeding CLs for individual stocks should be greater than 95% (ICES, 2012b).

3.1.2. Reference points in the NEAC area

River-specific CLs have been derived for salmon stocks in some countries in the NEAC area (France, Ireland, UK (England & Wales) and Norway). An interim approach has been developed for estimating national CLs for countries that cannot provide one based upon river-specific estimates. This approach is based on the establishment of pseudo stock–recruitment relationships for national salmon stocks (Potter *et al.*, 2004).

The NEAC-PFA run reconstruction model (see below) provides a means of relating estimates of the numbers of recruits to the numbers of spawners. The numbers of 1SW and MSW spawners are converted into numbers of eggs deposited using the proportion of female fish in each age class and the average number of eggs produced per female. The egg deposition in year 'n' is assumed to contribute to the recruitment in years "n+3" to "n+8" in proportion to the numbers of smolts produced of ages 1 to 6 years respectively. These proportions are then used to estimate the 'lagged egg deposition' contributing to the recruitment of maturing and non-maturing 1SW fish in the appropriate years. The plots of lagged eggs (stock) against the 1SW adults in the sea (recruits) are presented as 'pseudo stock–recruitment' relationships for each homewater country or region that is unable to provide river-specific CLs. In countries where with more than one region, the analysis is carried out for each region separately and the resulting estimates are summed to provide a national figure.

As noted previously, ICES currently define the CL for salmon as the stock size that will result in the maximum sustainable yield (MSY) in the long term. However, it is not straightforward to estimate this point on the stock-recruitment relationships established by the national PFA run-reconstruction models, as the replacement line (i.e. the line on which 'stock' equals 'recruits') is not known for these relationships. This is because the stock is expressed as eggs, while the recruits are expressed as adult salmon. To address this, WGNAS has developed a method for setting biological reference points from the national/ regional pseudo stock-recruitment datasets (ICES, 2001). This model assumes that there is a critical spawning stock level below which recruitment decreases linearly towards zero and above which recruitment remains constant. The position of this critical stock level is determined by searching for the stock value that provides the line of best fit for the stock and recruitment data provided by the PFA run-reconstruction model as determined by the residual sum of squares. This point is a proxy for Sim and is therefore defined as the CL for the stock, and is indicated by the inflection point in the hockey-stick relationship (e.g. see example at Figure 3.1.2.1).

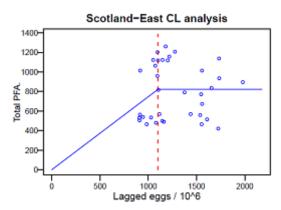


Figure 3.1.2.1. Pseudo stock-recruitment relationship for UK (Scotland) eastern region (from IC-ES, 2013).

Where river-specific estimates of CLs have been derived for all the rivers in a country or region, these are aggregated to provide national estimates. For countries where the development of river-specific CLs has not been completed, the method described above has been used (see example in Table 3.1.2.1, from ICES, 2013). The estimated national CLs are then summed to provide aggregate CLs for the northern and southern NEAC stock complexes (Table 3.1.2.1).

The CLs have also been used to estimate the spawning escapement reserves (SERs). These represent the CLs increased to take account of natural mortality between the recruitment date, 1st January, and the return to homewaters for maturing and non-maturing 1SW salmon from the northern NEAC and southern NEAC stock complexes (Table 3.1.2.1).

Table 3.1.2.1. Conservation limits (CLs) for NEAC countries and stock complexes estimated from river-specific values, where available, or the national PFA run-reconstruction model. Spawner escapement reserves (SERs) are also included for each stock complex (ICES 2016).

	National M	oderCLS	River Specific CLs		Conservation limit used		SER		
Northern Europe	1SW	MSW	1SW	MSW	1SW	MSW	1SW	MSW	
Finland			14,271	9,562	14,271	9,562	17,336	16,386	
Icland (north & east)	5,854	1,678			5,854	1,678	7,218	2,876	
Norway			60,614	72,747	60,614	72,747	77,009	120,991	
Russia	62,752	34,506			62,752	34,506	79,785	61,997	
Sweden			2,099	2,583	2,099	2,583	2,707	4,492	
			Stock Comp	lex	145,590	121,075	184,055	206,742	
	N-GIM		•		Occurrentia	- Barlans and		D	
	National M		River Spec	cific CLs	Conservation		SE		
	National M	odel CLs MSW	•		Conservation 1SW	n limit used MSW	SE 1SW	R MSW	
Southern Europe			River Spec	cific CLs					
Southern Europe			River Spec	cific CLs					
•			River Spec	ific CLs MSW	1SW	MSW	1SW	MSW	
France	1SW	MSW	River Spec	ific CLs MSW	1SW 17,400	MSW 5,100	1SW 22,440	MSW 9,419	
France Icland (south & west)	1SW	MSW	River Spec 1SW 17,400	5,100	1SW 17,400 17,790	5,100 1,171	1SW 22,440 21,935	MSW 9,419 2,006	
France Icland (south & west) Ireland	1SW	MSW		5,100 46,943	1SW 17,400 17,790 211,471	5,100 1,171 46,943	1SW 22,440 21,935 268,672	MSW 9,419 2,006 78,075	
France Icland (south & west) Ireland UK (E & W)	1SW	MSW	River Spec 1SW 17,400 211,471 53,988	5,100 46,943 29,918	1SW 17,400 17,790 211,471 53,988	5,100 1,171 46,943 29,918	1SW 22,440 21,935 268,672 68,591	9,419 2,006 78,075 51,271	

Table 3.2.2.1. Conservation limit options for NEAC stock groups

WGNAS considers the current CL and SER levels may be less appropriate to evaluating the historical status of stocks (e.g. pre-1985), that in many cases have been estimated with less precision.

3.1.3. Reference points in the NAC area

In many regions of North America, the CLs are calculated as the number of spawners required to fully seed the wetted area of the river. The methods and values used to derive the egg and spawner conservation requirements for Atlantic Canada are documented in O'Connell et al. (1997). CLs have generally been derived using freshwater production dynamics translated to adult returns to estimate the spawning stock for maximum sustainable yield (MSY). Data were available for a limited number of stocks and these values were transported to the remaining rivers using information on habitat area and the age composition of the spawners. A similar procedure was used to determine the CLs for rivers in the USA (ICES, 1995). In Québec, adult-toadult stock-recruitment relationships for six rivers were used to define the CLs for the other rivers (Caron et al., 1999). The definition of conservation in Canada varies by region and in some areas, historically the values used were equivalent to maximizing /optimizing freshwater production. These are used in Canada as limit reference points and they do not correspond to MSY values. Reference points for Atlantic salmon are currently being reviewed for conformity with the Precautionary Approach policy in Canada. Revised reference points are expected to be developed.

The NAC conservation requirements for 2SW salmon (only these are required in developing catch options for the West Greenland fishery) are summarised in Table 3.1.3.1 (from ICES, 2013). These are calculated from the adult age structure within the different regions and total 123 349 2SW salmon for Canada and 29 199 2SW salmon for the USA, for a combined total of 152 548.

Country and Comission Area	Stock Area	2SW spawne requiremen	
	Labrador	34 746	
	Newfoundland	4022	
	Gulf of St Lawrence	30 430	
	Québec	29 446	
	Scotia-Fundy	24 705	
Canada Total		123 349	
USA		29 199	
North American Total		152 548	

 Table 3.1.3.1. 2SW Conservation limits (CLs) for the six regions in the NAC area estimated from river-specific values.

3.2. Estimating PFA

Estimates of PFA are derived by run-reconstruction methods. These work back in time from estimates of abundance in homewaters to earlier periods of the salmon's life cycle by adding in catches at appropriate times and adjusting for survival. The run-reconstruction approach was first presented at ICES in 1992 and was subsequently adopted for stocks on both sides of the Atlantic (Rago *et al.*, 1993a; Potter and Dunkley, 1993; Potter *et al.*, 1998; 2004). The main advantage of backwards-running, run-reconstruction models over alternative forward-running approaches is that more extensive data are available on adult returns (e.g. traps, counters and catch data) than on freshwater production of juveniles. In addition, rates of natural mortality (M) were thought to be lower and more stable for large salmon after their first winter in the sea than during the post-smolt phase (Potter *et al.*, 2003).

The models used to estimate PFA take the generalised form:

$$PFA = Nh * \exp(Mt_h) + \sum_i C_i * \exp(Mt_i)$$

Where: Nh is the number of adult fish returning to homewaters, Ci the catch of fish from the stock in each interception fishery i (operating before the fish return to homewaters), M the monthly instantaneous rate of natural mortality of salmon in the sea after the first sea-winter, ti the time in months between the PFA date and the midpoint of fishery i, and th is the time in months between the PFA date and the midpoint of the return of fish to homewaters. Coastal catches are also added to the estimate where appropriate.

3.2.1. NEAC area run reconstruction model

The original model used to estimate the PFA of salmon from countries in the NEAC area was described by Potter *et al.* (2004); modifications have been described in subsequent WGNAS reports. PFA in the NEAC area is defined as the number of 1SW recruits on January 1st in the first sea winter. As there are relatively few fish of sea age three or more in most stocks, the model caters for two age groups, 1SW and MSW, the latter including all fish of sea age two or more that are treated as a single cohort. The model is therefore based on the annual catch in numbers of 1SW and MSW salmon in each country. These are raised to take account of minimum and maximum estimates of non-reported catches and exploitation rates of these two sea-age groups.

Thus, for each country (or region) c in year y, the total number of fish of sea age a caught in homewater fisheries ($Ch_{a,y,c}$) is calculated by dividing the declared catch ($Cd_{a,y,c}$) by the non-reporting rate (1 - $U_{a,y,c}$):

$$Ch_{a,y,c} = Cd_{a,y,c} / (1 - U_{a,y,c})$$

where: $U_{a,y,c}$ is the estimated proportion of the total catch that is unreported or discarded. The number of fish returning to homewaters (Nh_{a,y,c}) is estimated by dividing the total homewater catch by the exploitation rate (H_{a,y,c}):

$$Nh_{a,y,c} = Ch_{a,y,c} / H_{a,y,c}$$

As the model provides estimates of total returns and total catch (including non-catch fishing mortality), it is then also possible to estimate the spawner escapement (Ns_{a,y,c}):

 $Ns_{a,y,c} = Nh_{a,y,c} - Ch_{a,y,c}$

Total catches in the Faroese (Cf_{a,y}) and West Greenland (Cg_{a,y}) fisheries are similarly calculated by correcting the declared catches for non-reporting, but they are not raised for the exploitation rate, because the uncaught fish are accounted for from the returns to homewaters. The West Greenland fishery only exploits salmon that would otherwise mature as MSW fish, although the majority are 1SW fish in the summer that they are caught; for the purpose of the model, all are classed as 1SW. The Faroese fishery exploits predominantly MSW salmon, but also a small number of 1SW fish, 78% of which have been estimated to be maturing (ICES, 1994). Over the past two decades, a substantial proportion of the fish caught in the Faroese fishery have been escapees from salmon farms, and these are discounted from the assessment of wild

stocks on the basis of data from Hansen *et al.* (1999). The incidence of farm escapees in the West Greenland catch is thought to be <1.5% (Hansen *et al.*, 1997), so this portion is ignored in the model. The total estimated catches of wild fish in both distantwater fisheries are assigned to the PFA for different countries on the basis of historic tagging studies (Potter, 1996).

The returns to homewaters and catches in the distant-water fisheries of 1SW and MSW salmon are then raised to take account of the marine mortality between January 1st in the first sea winter (the PFA date) and the mid-point of the period over which the respective national fisheries operate. WGNAS determined a natural mortality value of 0.03 per month to be the most appropriate (ICES, 2002) and a range 0.02 to 0.04 is applied within the model in a Monte Carlo simulation. Thus, the PFA of maturing 1SW fish (PFAm), survivors of which will return to homewaters as 1SW adults, is:

```
PFAm_{y,c} = Nh_{1,y,c} * exp(Mt_{h,1,c}) + 0.78 * Cf_{1,y} * w_y * pf_{1,c} * exp(Mt_{f,1,c})
```

and the PFA of non-maturing 1SW fish (PFAn), survivors of which will return to homewaters as MSW adults, is:

$$\begin{aligned} & \text{PFAn}_{y,c} = \text{Nh}_{2,y+1,c} * \exp(\text{Mt}_{h,2,c}) + \text{Cg}_{1,y} * \text{pg}_{1,c} * \exp(\text{Mt}_{g,1,c}) \\ & + 0.22 * \text{Cf}_{1,y} * w_y * \text{pf}_{1,c} * \exp(\text{Mt}_{f,1,c}) + \text{Cf}_{2,y+1} * w_{y+1} * \text{pf}_{2,c} * \exp(\text{Mt}_{f,2,c}) \end{aligned}$$

where indices y and c represent year and country/region, indices 1 and 2 the 1SW and MSW sea age groups, w is the proportion of the Faroese catch that is of wild origin, pf and pg are the proportion of the catches in the Faroese and West Greenland fisheries originating in each country (as indexed), and th, tr and tg are the times in months between the PFA date and the midpoints of the homewater fisheries, the Faroese fishery, and the West Greenland fishery, respectively, for the year classes and country/region as indexed.

Total 1SW recruitment for the NEAC area in year y is therefore the sum of the maturing 1SW and non-maturing 1SW recruitments for that year for all countries:

$$PFA_{y} = \sum_{c} PFAm_{y,c} + \sum_{c} PFAn_{y,c}$$

The non-reporting rates, exploitation rates, natural mortality, and migration times in the above equations cannot be estimated precisely, so national experts provide minimum and maximum values based upon best available knowledge that are considered likely to be centred on the true values (ICES, 2003). A Monte Carlo simulation (12 000 trials) is used to estimate confidence intervals on the stock estimates.

Where appropriate to the provision of management advice, the national outputs from the model are combined into stock complexes, such as those for southern and northern NEAC (ICES, 2002). The confidence limits for these combined estimates are derived from the sum of the national variances obtained from the MCS (the covariances are assumed to be small). This model has provided time-series of PFA estimates for NEAC salmon stocks from 1971 to the present.

The model was initially run using 'Crystal Ball' (CB) in Excel (Decisioneering, 1996). However, an updated version of the model which runs in the 'R' programming language (R Development Core Team, 2007) was developed in 2011 (ICES, 2011). This provided a more flexible platform for the further development of the model and to allow its integration with the Bayesian forecast model for the development of catch options (see below). In 2012, the outputs of the CB and 'R' models were compared to examine the approaches taken and to validate the outputs (ICES, 2012a). Since 2013, the run-reconstruction analysis has been completed by WGNAS using the 'R' programme (ICES, 2013). This has also enabled additional sources of uncertainty to be incorporated into the modelling approach (ICES, 2013).

The full set of country-specific data inputs, as used in the most recent assessment (ICES, 2016) is provided at Appendix 3. The 'R' code used for running the model and the additional data input file required to run the model, are available on the ICES WGNAS SharePoint site.

3.2.2. NAC area run reconstruction model

The run–reconstruction model developed by Rago *et al.* (1993a) and described in previous WGNAS reports (ICES, 2008; 2009a) and in the primary literature (Chaput *et al.*, 2005) is used to estimate returns and spawners by size (small salmon, large salmon) and sea age group (2SW salmon) to the six geographic regions of NAC from 1971 to the present. The model takes the form:

 $PFA_{year(i)} = [NR2_{year(i+1)} * e^{MX1} + NC2_{year(i+1)}] * e^{MX10} + NC1_{year(i)} + NG1_{year(i)}$

where: NR2_{year(i+1)} is the sum of 2SW returns to six regions of North America in year i + 1, NC2_{year(i+1)} is the catch of 2SW salmon in Newfoundland and Labrador commercial fisheries in year i + 1, NC1_{year(i)} is the catch of 1SW non-maturing salmon in Newfoundland and Labrador commercial fisheries in year i, NG1_{year(i)} is the catch of 1SW non-maturing salmon of North American origin in the Greenland fishery in year i, and M is the monthly instantaneous natural mortality of 0.03.

The reconstruction begins with the estimation of returns of 2SW salmon in year i + 1 to six regions in eastern North America: Labrador, Newfoundland, Québec, Gulf, Scotia-Fundy, and USA. For the four southern regions, the regional returns include the harvest in the coastal commercial fisheries but this is not the case for Newfoundland and Labrador. For Labrador, the returns to rivers are estimated from the commercial harvest factored by an exploitation rate. The harvest of 2SW salmon in the Newfoundland and Labrador mixed-stock fisheries in year i + 1 is added to the sum of the returns to the six regions (prorated backward for one month of natural mortality - equates to 1st June of year i + 1) to produce the returns to North America. Finally, the harvests of North American origin salmon in the Greenland fisheries in year i and the harvest of non-maturing 1SW salmon in the Newfoundland and Labrador commercial fisheries in year i are added to the prorated returns to North America (ten months between abundance at Greenland on 1st August year i and North America on 1st June year i + 1) to produce the pre-fishery abundance of non-maturing 1SW salmon of North American origin. An instantaneous natural mortality rate of 0.03 per month is assumed for salmon in the second year at sea for all years (ICES, 2002). Adjustments to the input data resulting from reductions and subsequent closures of commercial fisheries in North America are summarized by Friedland et al. (2003).

Following earlier WGNAS recommendations (ICES, 2008), the run–reconstruction model since 2009 has been developed using Monte Carlo simulation (OpenBUGS; <u>http://mathstat.helsinki.fi/openbugs/</u>; Lunn *et al.*, 2000). This is similar to the approach applied for the NEAC area.

The PFA of the non-maturing component of 1SW fish, destined to be 2SW returns (excluding 3SW and previous spawners) is the estimated number of salmon in the North Atlantic on August 1st of the second summer at sea. As this requires estimates of 2SW returns to rivers, there is always a lag in providing this figure (PFA estimates for year n require 2SW returns to rivers in North America in year n + 1).

The full set of data inputs, as used in the most recent assessment (ICES, 2015) is provided at Appendix 4. The 'R' code used for running the model is available on the WGNAS SharePoint site.

3.2.3. Instantaneous natural mortality rate (M)

The natural mortality rate for salmon after they recruit to the distant water fisheries has been the subject of much discussion. WGNAS originally used a value of 0.01 per month, based upon Doubleday *et al.* (1979), but this was modified to 0.03 per month following a detailed review as part of the EU SALMODEL project (Crozier *et al.*, 2003; ICES, 2002) on the basis of inverse-weight and maturity-schedule models. The rate is assumed to have been constant over the time-series. While mortality may be expected to vary among years and may also be different for maturing and non-maturing 1SW recruits, WGNAS has not had data on which to base the use of different values, or values that change over time. The assumption is, therefore, that the mortality of adult fish after the first winter at sea has not changed and that all the variability of marine mortality has occurred at the post-smolt stage. Efforts are continuing to explore levels of mortality and to better partition this between different stages of the life cycle. The issue was also subject to further investigation within the EU ECOKNOWS project and Bayesian modelling may provide alternative approaches in future.

3.3. PFA forecast models

3.3.1. Introduction

The provision of quantitative catch advice for the distant water fisheries requires estimates of abundance before the fisheries take place. While there has been some use of in-season surveys in the management of these fisheries (NASCO, 2001), such methods are considered too impractical and costly to implement on a widespread scale (Potter *et al.*, 2004). Models have therefore been developed by WGNAS which relate abundance estimates obtained at other life stages to the PFA. The objective has been to account for this relationship in terms of biological or environmental factors that affect natural mortality, and to use this to forecast future stock levels.

An initial PFA forecast model for North American stocks (Rago *et al.*, 1993b) utilised indices of thermal habitat in relation to historically observed PFA (from the runreconstruction model) to predict future PFA. Similar approaches were explored by Crozier *et al.* (2003) for NEAC stocks. However, while statistically significant temperature indices could be constructed, the relationships were not always consistent or intuitively correct. Alternative approaches were therefore explored for NEAC; these are described by Potter *et al.* (2004). More recently work by the ICES Study Group on Salmon Stock Assessment and Forecasting (SGSSAFE) has, however, resulted in the development of Bayesian forecast models for both NAC and NEAC (ICES, 2009a; 2011; Chaput, 2012).

In the latest models, PFA dynamics by complex are modelled using the estimates of adult spawners, adjusted to the number of eggs per fish based on life-history charac-

teristics of the age groups within each region of the stock complexes (ICES, 2011; Chaput, 2012). The spawner to PFA dynamic is modelled as:

 $PFA_y = e^{\alpha_y} LE_y e^{\varepsilon}$

where: α_y is the productivity parameter from eggs (×1000) to PFA (number of fish) for PFA year *y* (on a log-scale), LE_y the estimated lagged eggs (×1000) corresponding to the PFA cohort in year *y*, and the progress of α_y is modelled as $a_{y+1} = a_y + \varepsilon$, with $\varepsilon \sim N(0, \sigma^2)$.

Productivity is modelled as an integration of survival in freshwater and during the first year at sea. An important assumption is the absence of heritability of age at maturity, i.e. all eggs are considered equivalent regardless of the age of the spawners. Lagged eggs refer to the adjustment of the egg depositions to correspond to the expected age at smoltification. Spawners in year 'n' contribute to recruitment in years 'n+3' to 'n+8' depending upon the relative proportions of one to six year-old smolts that they produce. For example, spawners in year 'n' produce eggs that hatch in year 'n+1' and may produce one year-old smolts in year 'n+2', which would become 1SW recruits in year 'n+3'. Any two year-old smolts from the same spawners would produce 1SW recruits in year 'n+4', etc.

At the stock complex level, lagged eggs are the sum of the eggs from the spawners in year y - (s + 2) weighted by the proportion of the smolts produced at age s in region k summed over regions in the complex. Two years are added to the smolt age, for the spawning year and smolt migration year, to lag the eggs to the corresponding year of PFA:

$$LE_y = \sum_k \sum_s Eggs_{y-(s+2),k} \circ \operatorname{prop}_{s,k}$$

3.3.2. NEAC PFA Forecast model

A forecast model to estimate PFA for all four NEAC stock complexes has been developed in a Bayesian framework by the Study Group on Salmon Stock Assessment and Forecasting (SGSSAFE). The model was originally reported to WGNAS in 2009 (ICES, 2009a), but was subsequently refined and has been in use by WGNAS in its present form since 2011 (ICES, 2011). The models for the northern and southern NEAC stock complexes have exactly the same structure and are run independently. A Directed Acyclical Graph (DAG) for the models is provided in Figure 3.3.2.1. The model considers both the maturing *PFA* (denoted *PFAm*) and the non-maturing PFA (denoted *PFAnm*). The full code used for running the model is available on the WGNAS Share-Point site.

A disaggregated version of the Bayesian model has since been developed using the same structure to provide forecasts at a country level, for all countries in both southern and northern NEAC model implementations (ICES, 2013). In these, countries are linked hierarchically only through the variance on the productivity parameter "a". There is no modelling linkage between the northern and southern complexes.

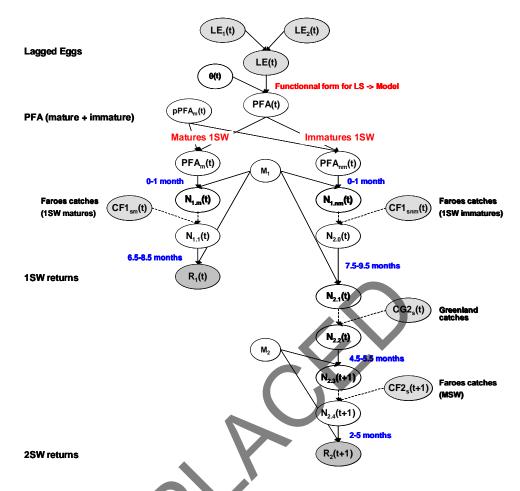


Figure 3.3.2.1. Directed Acyclical Graph (DAG) of the structure of the combined sea age model for the southern NEAC and northern NEAC forecast models. Ellipses in grey are observations (or pseudo-observations) derived from sampling programmes or from submodels (runreconstruction).

The *PFA* is modelled using the summation of lagged eggs from 1SW and MSW fish (*LE*) for each year *t* and an exponential productivity parameter (*a*).

 $PFAt = LEt^* exp(at)$

The productivity parameter (*a*) is the proportionality coefficient between lagged eggs and PFA. This is forecasted one year at a time (a_{t+1}) in an auto correlated random walk, using the previous year's value (*a*) as the mean value in a normal distribution, with a common variance for the time-series of *a*.

 $at+1=at+\varepsilon$; $\varepsilon \sim N(0, a.\sigma^2)$

The maturing *PFA* (denoted *PFAm*) and the non-maturing *PFA* (denoted *PFAnm*) recruitment streams are subsequently calculated from the proportion of *PFA* maturing (*p.PFAm*) for each year *t. p.PFAm* is forecast as an autocorrelated value from a normal distribution based on a logit scale, using the previous year's value as the mean and a common variance across the time-series of *p.PFAm*.

 $logit.p.PFAmt+1 \sim N(logit.p.PFAmt, p.\sigma2)$ logit.p.PFAmt = logit (p.PFAmt) Uncertainties in the lagged eggs were accounted for by assuming that the lagged eggs of 1SW and MSW fish were normally distributed with means and standard deviations derived from the Monte-Carlo run-reconstruction at the scale of the stock complex. The uncertainties in the maturing and non-maturing PFA returns are derived in the Bayesian forecast models through the pseudo-observation method proposed by Michielsens *et al.* (2008), as used in the NAC model.

The natural mortality in the post-PFA time point was assumed constant among years, centred on an instantaneous rate value of 0.03 per month with a 95% confidence interval range of 0.02 to 0.04.

Catches of salmon at sea in the West Greenland fishery (as 1SW non-maturing salmon) and at Faroes (as 1SW maturing and MSW salmon) were introduced as covariates and incorporated directly within the inference and forecast structure of the model. For southern NEAC, the data are available for a time-series of lagged eggs and returns commencing in 1978. Although the return estimates to southern NEAC begin in 1971, the lagged eggs are only available from 1978 due to the smolt age distributions (one to five years). For northern NEAC, data are available for a shorter time-series. Return and spawner estimates begin in 1983, but due to the smolt age distributions (one to six years), the lagged eggs are only available from 1991 onward. The models are fitted and forecasts derived in a consistent Bayesian framework.

The model provides forecasts for maturing and non-maturing stocks for both southern and northern NEAC complexes (and countries) for five years. Risks are defined each year as the posterior probability that the PFA would be greater than or equal to the age and stock complex/ country specific Spawner Escapement Reserves (SERs), under the scenario of no exploitation.

The country disaggregated version of the Bayesian NEAC inference and forecast model incorporates country specific catch proportions at Faroes, lagged eggs and returns of maturing and non-maturing components. Model structure and operation is as described above, incorporating country and year indexing. Linkage between countries in the model is through the common variance parameter associated with the productivity parameter (*a*) (the proportionality coefficient between lagged eggs and PFA), which is forecast forward and used along with the forecast proportion maturing to estimate the future maturing and non-maturing PFAs. The evolution of *a* is independent between countries with the exception of its associated variance. Evolution of the proportion maturing (*p.PFAm*) is also independent for each country, as is its variance.

3.3.3. NAC PFA Forecast model

WGNAS (ICES, 2009; 2012a) developed forecasts of the pre-fishery abundance for the non-maturing 1SW salmon (PFA) using a Bayesian framework that incorporates the estimates of lagged spawners and works through the fisheries at sea to determine the corresponding returns of 2SW salmon, conditioned by fisheries removals and natural mortality at sea. This model considers regionally-disaggregated lagged spawners and returns of 2SW salmon for the six regions of North America. The model is summarised in the Directed Acyclical Graph in Figure 3.3.1.1. The year is identified by the i index. The full code used for running the model is available on the WGNAS Share-Point site.

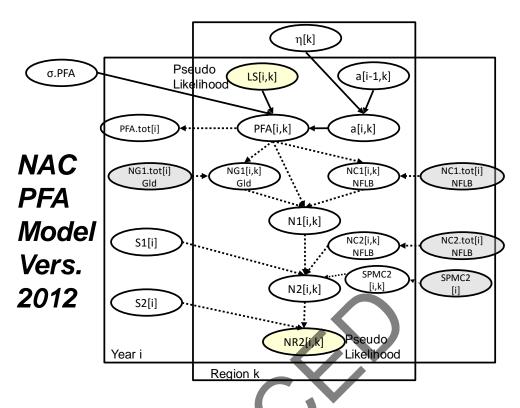


Figure 3.3.1.1. Directed Acyclical Graph (DAG) of the structure of the region disaggregated forecast model for 2SW salmon of North American origin. Ellipses in grey are observations (or pseudo-observations) derived from sampling programmes or from submodels (run-reconstruction).

Lagged spawners LS_{i,k} represent the sum of smolt age adjusted annual spawners by region (k) that would be expected to contribute to the recruitment at sea prior to the fisheries (PFA) for year i. LS_{i,k} are not directly observed but are estimated from the run-reconstruction submodel used to estimates returns and spawners to each of the six regions.

The probability distributions of LS (and returns of 2SW) by region are used as likelihood functions expressing comparative degrees of belief given the data and a probability model not explicitly specified in the current model. The probability distributions were drawn from the Monte Carlo simulations and assumed to be normal with known mean (LS.m) and precision (1/variance) (tau.LS). The use of this distribution as a likelihood function is equivalent to assuming a pseudo-observation equal to LS.m issuing from a sampling distribution with mean and precision equal to LS and tau.LS (Michielsens *et al.*, 2008).

 $LS.m_{i,k} \sim N (LS_{i,k}, tau.LS_{i,k})$

The LS.mi,k (mean) and tau.LS_{i,k} (precision) were derived assuming the lagged spawner values issued from a normal distribution characterized by the 95% confidence interval range statistics retained from the Monte Carlo simulations of returns.

Similarly, the returns of 2SW salmon to the six regions (NR2_{i,k}) are not directly observed, but estimated from the run-reconstruction model. The probability distributions were assumed to be normal with known mean NR2.m and variance tau.NR2. As with the LS variable, the NR2 were treated as pseudo-observations equal to NR2.m issuing from normal sampling distributions with means and variances equal to NR2 and tau.NR2. NR2. $m_{i,k} \sim N$ (NR2 $_{i,k}$, tau.NR2 $_{i,k}$)

In between the lagged spawners and returns as 2SW salmon, the catches in the various sea fisheries and conditioning for natural mortality as the fish move from the time of the PFA to homewaters, are incorporated (Figure 3.3.1.1). The catches in the commercial fisheries of West Greenland and the Newfoundland and Labrador commercial and coastal fisheries (NG1.tot, NC1.tot and NC2.tot) are not directly observed, but estimated with error. The catches are converted to numbers of fish of 1SW non-maturing and 2SW fish based on the characteristics of the fish in the catch. Their (prior) probability distributions are obtained from catch statistics according to a formal structure included in the model.

Catches of large salmon (assumed to be 2SW salmon) from the St Pierre & Miquelon fisheries (SPMC) are also included in the model as point estimates.

The natural mortality in the post-PFA time point was assumed constant between years, centred on an instantaneous rate value of 0.03 per month (95% confidence interval range of 0.02 to 0.04).

For the NAC 2SW component, the model is fitted to an historical dataseries of lagged eggs starting from 1978. Although the return and spawner estimates for NAC begin in 1971, the lagged eggs are only available from 1978 due to the smolt age distributions (one to six years).

The years are modelled independently conditionally on the lagged spawners and yearly productivity parameters. The lagged spawners to PFA ratios (productivity) are modelled dynamically, i.e. assuming they are sequentially dependent within a region and attempts to take into account the most significant sources of uncertainty. The DAG for the model is shown in Figure 2.

PFA is assumed to be proportional to lagged-spawners (LS), with i.i.d. lognormal errors, and is modelled separately for each region (k = 6). The first year in the timeseries (t) is 1978 for lagged spawners (due to the range of smolt ages 1 to 6 for NAC and the start of the spawner time-series in 1970) and the last year of lagged spawner data is for the 2017 PFA year. The PFA can be modelled for 1978 to 2013 (the last PFA year for which returns of 2SW salmon have been estimated back to rivers in 2014).

$$\begin{aligned} PFA_{i,k} &= LogN(\overline{PFA_{i,k}}, \sigma.PFA^2) \\ \overline{PFA_{i,k}} &= \log(LS_{i,k}) + a_{i,k} + \varepsilon_{i,k} \\ \varepsilon_{i,k} & \stackrel{iid}{\sim} N(0, \sigma.PFA^2) \end{aligned}$$

The total PFA is calculated as the sum of the regional PFA's (k = 6). The proportion of the total PFA in each region is calculated directly as:

 $p.PFA_{i,k} = PFA_{i,k} / PFA.tot_i$

A non-informative prior is assumed for σ .*PFA*² (1/ σ .*PFA*² ~ *gamma*(0.01, 0.01)

The proportionality coefficient (log) $a_{i,k}$ between LS_{i,k} and PFA_{i,k} for each region is modelled dynamically as a random walk with a year and region residual variation ($\eta_{i,k}$) assumed multivariate normal (MVN). The variance covariance matrix (Σ) allows for correlations among regional productivity values reflecting that the fish share a common marine environment during part of their life cycle and that there are regional specificities in the evolution of the freshwater or the marine coastal environment.

$$a_{i+1,k} = a_{i,k} + \eta_{i,k}$$

$$\eta_{i,k} \sim MVN(0, \Sigma)$$

The common yearly evolution of a is the mean of annual a across regions:

 $a.y_i \leq mean(a_{i,k})$

This parameterization of the covariance of the proportionality coefficient differs from an earlier version of the model for which the proportionality (log) coefficient $a_{i,k}$ between LS_{i,k} and PFA_{i,k} for each region dynamically as a random walk model with the addition of a regionally common annually varying parameter (e.y_i).

$$a_{i+1,k} = a_{i,k} + e.y_{i+1} + \omega_{i,k} \quad \text{with} \quad \omega_{i,k} \stackrel{i.i.d}{\sim} N(0, a\sigma_k^2)$$
$$e.y_i \stackrel{i.i.d}{\sim} N(0, y.\sigma^2)$$

The correlation matrix of *a* among the regions is calculated from the covariance matrix:

- 1) the precision matrix is inverted to produce the covariance matrix;
- 2) the covariance matrix is transformed to the correlation matrix.

The positive-definite matrix (T, the precision matrix) is inverted:

covariance matrix <- solve(T)

correlation matrix <- cov2cor(b)

The dynamic component of the model requires initialization for the first year (i = 1978) and an uninformative prior is assumed:

$$a_{1,k} \stackrel{i.i.d}{\sim} N(0, 100)$$

The models are fitted and forecasts derived in a consistent Bayesian framework under the OpenBUGS 3.0.3 software (<u>http://mathstat.helsinki.fi/openbugs/</u>; Lunn *et al.*, 2000).

3.3.4. Summary of NAC and NEAC forecast models

The data inputs and models currently used by WGNAS for forecasting and provision of catch advice differ between the Commission areas; outline details are summarised in the text table below.

	FORECAST MODELS	
	NAC	NEAC
Data inputs		
Time period of data	1978 on	1978 on for southern NEAC 1991 on for northern NEAC
Spatial aggregation	Separately for six regions of North America	By southern and northern stock complexes & NEAC countries
Age components	2SW salmon component only	1SW and MSW age components
Spawners	Lagged spawners by region for 2SW salmon only	Lagged eggs by sea age component for the southern and northern complexes/ country
Returns	Returns by region of 2SW salmon only	Returns of 1SW and MSW age components by stock complex / country
Model structure		
Spatial aggregation	Spawners and returns of 2SW salmon for six regions	Spawners and returns for two sea age components for the southern and northern NEAC complexes / countries
Dynamic function	Random walk dynamic	Random walk dynamic
	Region-specific recruitment rates linked with an annual recruitment rate variable	Sea age specific recruitment rates linked with a probability of maturing variable
Latent variables of interest	PFA 1SW non-maturing Recruitment rate by region and year	PFA 1SW maturing and PFA 1SW non-maturing by stock complex/ country Recruitment rate by sea age
		component and the probability of maturing variable
Forecast years	Four years	Five years; i.e. the present year -1, the present year, and the next three years
		(y-1 is a forecast, as the MSW stock component is yet to return).

3.4. The development of a risk analysis framework for catch advice

3.4.1. Introduction

The provision of catch advice in a risk framework involves incorporating the uncertainty in all the factors used to develop the catch options (ICES, 2002). The ranges in the uncertainties of all the factors will result in assessments of differing levels of precision. The analysis of risk involves four steps:

- 1) identifying the sources of uncertainty;
- 2) describing the precision or imprecision of the assessment;
- 3) defining a management strategy; and
- 4) evaluating the probability of an event (either desirable or undesirable) resulting from the fishery action.

The uncertainties have been identified and quantified in the assessment of PFA for salmon stocks in both the NAC and NEAC areas. NASCO's strategy for the management of salmon fisheries is based upon the principle of ensuring that stocks are above CLs (defined in terms of spawner escapement or egg deposition) with a high probability. The undesirable event to be avoided is that the spawning escapement after the fisheries will be below the CLs.

3.4.2. Catch advice and risk analysis framework for the West Greenland fishery

A risk framework for the provision of catch advice for the West Greenland fishery has been applied since 2003 (ICES, 2003) and has been subject to a number of subsequent updates. The current procedure is outlined below. This involves estimating the uncertainty in meeting defined management objectives at different levels of catch (catch options). The risk framework has been formally accepted by NASCO.

Two stock complexes are of relevance to the management of the West Greenland fishery; non-maturing 1SW fish from North America and non-maturing 1SW fish from southern NEAC. The risk assessments for the two stock complexes are developed in parallel and then combined at the end of the process into a single summary plot or catch options table. The risk analysis proceeds as illustrated in the flowchart in Figure 3.4.2.1).

The primary inputs to the risk analysis for the complex at West Greenland are:

- PFA forecast for the year of the fishery; PFANA and PFANEAC;
- Harvest level being considered (t of salmon);
- Conservation spawning limits or alternate management objectives; and
- The post-fishery returns to each region.

The risk analysis of catch options incorporates the following input parameter uncertainties: (i) the uncertainty of the pre-fishery abundance forecast, (ii) the uncertainty in the biological parameters used to translate catches (weight) into numbers of salmon, and (iii) the uncertainty in attaining the conservation requirements simultaneously in different regions.

The uncertainty in the PFA_{NA} and PFA_{NEAC} is accounted for in the forecast approach described above. The number of 1SW non-maturing fish of North American and European origin in a given catch (t) is conditioned by the continent of origin of the fish (prop_{NA}, prop_E), by the average weight of the fish in the fishery (Wt_{Allages}), and by the proportion 1SW non-maturing fish in the respective continent of origin catches. These parameters define how many fish originating in North America and Europe are expected in the fishery harvests. For a level of fishery under consideration, the weight of the catch is converted to number of fish of each continent of origin using the following equation:

$$C1SW_{c} = \frac{t X propC}{ACF * (propNA X Wt1SW_{NA} + propE X Wt1SW_{z})}$$

where: *C1SWc* is the catch (number of fish) of 1SW salmon originating in continent C (either North America or Europe), *t* is the fishery harvest at West Greenland in kg, *propC* is the proportion of the 1SW salmon harvest which originates from continent C, *Wt1SW*_{NA} and *Wt1SW*_E are the average weight (kg) in the fishery of a 1SW salmon of North American and European origin, respectively, and *ACF* is the age correction factor by weight for salmon in the fishery which are not at age 1SW.

Since these parameters are not known for the forecast years of interest, they are estimated from previous values. Thus, propNA (and propNEAC as 1 - propNA) are drawn randomly from observed values of the past five years taking account of uncertainty due to sample sizes. For the other parameters, it is assumed that the parameters for WtAllages and the proportion non-maturing 1SW in the catch by continent of origin could vary uniformly within the values observed in the past five years.

For a level of fishery under consideration, the weight of the catch is converted to fish of each continent of origin and subtracted from one of the simulated forecast values of PFA_{NA} and PFA_{NEAC}. The fish that escape the Greenland fishery are immediately discounted by the fixed sharing fraction (F_{na}) historically used in the negotiations of the West Greenland fishery. The sharing fraction chosen is the 40%:60% West Greenland:North America split. The same sharing arrangement has been used for NEAC stocks (ICES, 2003). [Any sharing fraction could be considered and incorporated at this stage of the risk assessment].

After the fishery, fish returning to homewaters are discounted for natural mortality from the time they leave West Greenland to the time they return to rivers. For North America this is a total of eleven months at a rate of M = 0.03 (equates to 28.1% mortality). For southern NEAC stocks this is a total of eight months at a rate of M = 0.03 (equates to 21% mortality). The fish that survive to North American homewaters are then distributed among the regions based on the regional proportions of lagged spawners for the last five years when estimates of spawners were available. The uncertainty in the regional proportions is characterised by drawing at random from a uniform distribution defined by the minimum and maximum regional ranges from the five years and calculating the average proportion for each of the six regions in North America.

nerica.

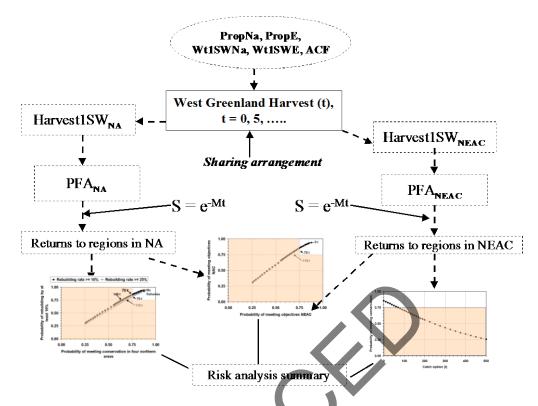


Figure 3.4.2.1. Flow chart summarising risk analysis for catch options at West Greenland using the PFANA and the PFANEAC predictions for the year of the fishery. Inputs with solid borders are considered known without error. Estimated inputs with observation error that is incorporated in the analysis have dashed borders. Solid arrows are functions that introduce or transfer without error whereas dashed arrows transfer errors through the components.

The final step in the risk analysis of the catch options involves combining the conservation requirement or alternate management objectives with the probability distribution of the returns to North America for different catch options. Estimated 2SW returns to each region are compared to the conservation objectives of Labrador, Newfoundland, Québec, and Gulf. Estimated returns for Scotia-Fundy are compared to the objective of achieving an increase of 25% relative to average returns of the base period, 1992–1996. For the USA, the management objective was revised in 2014 (ICES, 2014). Estimated returns for the USA are now compared to the objective of achieving 2SW adult returns of 4549 or greater. The advice to fisheries managers is presented as a probability plot (or table) of meeting or exceeding the objectives relative to increasing harvest levels at West Greenland.

ICES has adopted, a risk level of 75% of simultaneous attainment of seven management objectives (ICES, 2003) as part of an agreed management plan for the West Greenland fishery. The same level of risk aversion is applied for catch advice for homewater fisheries on the North American stock complex.

The catch advice for the West Greenland fishery is currently tabulated to show the probability of each management unit achieving its CL (or alternative reference level) individually and the probability of this being achieved by all management units simultaneously (i.e. in the same given year) (e.g. ICES, 2012a). This allows managers to evaluate both individual and simultaneous attainment levels in making their management decisions. Table 3.4.2.1 provides an example of catch options for West Greenland for the years 2012 to 2014 (ICES, 2012a). An updated catch options table

for 2015 to 2017 is provided in Section 5.3 of the 2015 Working Group report (see above).

The models currently used by WGNAS in developing catch advice are considered to provide consistent characterisation of the status and expectations for Atlantic salmon in the North Atlantic. Compared to previous models used by WGNAS prior to 2009, the Bayesian models provide more flexibility, are consistent with the emerging emphasis on such approaches in natural resource assessment, and can provide management advice consistent with the probability of achieving management objectives.

Table 3.4.2.1. Catch options tables for mixed-stock fishery at West Greenland by year of PFA, 2012to 2014.

option (t)	LAB	NELD						
0	LAB NFLD QC GULF		GULF	SF	USA	S-NEAC	ALL	
	0.45	0.86	0.71	0.50	0.15	0.89	0.98	0.06
10	0.42	0.84	0.67	0.48	0.14 0.88		0.98	0.05
20	0.40	0.83	0.63	0.45	0.13 0.87		0.98	0.05
30	0.38	0.81	0.59	0.42	0.12	0.85	0.98	0.04
40	0.36	0.78	0.54	0.40	0.12	0.83	0.98	0.04
50	0.34	0.76	0.50	0.38	0.11	0.81	0.98	0.03
60	0.32	0.73	0.46	0.36	0.10	0.79	0.98	0.03
70	0.30	0.70	0.42	0.33	0.09	0.77	0.98	0.03
80	0.28	0.67	0.39	0.31	0.08	0.74	0.98	0.03
90	0.26	0.64	0.35	0.29	0.08	0.72	0.97	0.02
100	0.24	0.60	0.32	0.27	0.07	0.68	0.97	0.02
2013 Catch	Probabi	lity of meet	ing or exce	eding regio	on-specific	manageme	ent objectives	
Option (t)	LAB	NFLD	QC	GULF	SF	USA	S-NEAC	ALL
0	0.48	0.78	0.73	0.50	0.25	0.75	0.95	0.08
10	0.46	0.76	0.70	0.48	0.24	0.73	0.95	0.07
20	0.44	0.75	0.67	0.46	0.23	0.72	0.95	0.06
30	0.42	0.73	0.63	0.44	0.22	0.70	0.95	0.06
40	0.41	0.70	0.60	0.42	0.21	0.68	0.95	0.06
50	0.39	0.68	0.56	0.40	0.20	0.66	0.94	0.05
60	0.37	0.65	0.53	0.38	0.19	0.64	0.94	0.05
70	0.35	0.63	0.50	0.36	0.18	0.62	0.94	0.05
80	0.33	0.60	0.47	0.34	0.17	0.59	0.94	0.04
90	0.31	0.57	0.44	0.32	0.16	0.57	0.94	0.04
100	0.30	0.54	0.41	0.31	0.15	0.55	0.94	0.04
2014 Catch	Probabi	lity of meet	ing or exce	eding regio	on-specific	managem	ent objectives	
Option (t)	LAB	NFLD	QC	GULF	SF	USA	S-NEAC	ALL
0	0.56	0.78	0.75	0.55	0.20	0.86	0.94	0.08
10	0.55	0.77	0.73	0.53	0.20	0.85	0.94	0.08
20	0.53	0.75	0.70	0.51	0.19	0.84	0.94	0.07
30	0.52	0.73	0.67	0.49	0.18	0.83	0.94	0.07
40	0.50	0.71	0.64	0.47	0.17	0.82	0.94	0.06
50	0.48	0.69	0.62	0.46	0.17	0.81	0.94	0.06
		0.67	0.59	0.44	0.16	0.79	0.94	0.06

2012 Catch	Probability of meeting or exceeding region-specific management objectives									
option (t)	LAB NFLD QC		GULF SF		USA	S-NEAC	ALL			
70	0.45	0.65	0.56	0.42	0.16	0.77	0.94	0.05		
80	0.43	0.63	0.54	0.41	0.15	0.76	0.94	0.05		
90	0.42	0.61	0.51	0.39	0.14	0.74	0.94	0.05		
100	0.40	0.59	0.49	0.38	0.14	0.72	0.94	0.05		

3.4.3. Catch advice and risk analysis framework for the Faroes fishery

3.4.3.1. Outline of the risk framework and management decisions required

There is currently no agreed framework for the provision of catch advice for the Faroes fishery adopted by NASCO. However, NASCO has asked ICES, for a number of years, to provide catch options or alternative management advice with an assessment of risks relative to the objective of exceeding stock conservation limits for salmon in the NEAC area. An initial risk framework that could be used to provide and evaluate catch options for the Faroes fishery was outlined by WGNAS in 2010 (ICES, 2010a). This was based on the method currently used to provide catch advice for the West Greenland fishery, which involves estimating the uncertainty in meeting defined management objectives at different catch levels (TAC options). The Faroes risk framework was developed further at subsequent WGNAS meetings (ICES, 2011; 2012a) and the current proposed procedure is outlined below.

A number of decisions are required by managers before full catch advice could be provided (ICES, 2011; 2012a), Specifically, ICES has indicated that NASCO would need to agree upon the following issues before the risk framework could be finalised:

- season (January to December or October to May) to which any TAC should apply;
- share arrangement for the Faroes fishery;
- choice of management units for NEAC stocks; and
- specification of management objectives.

In developing an indicative risk framework, WGNAS has made pragmatic choices regarding these issues:

Faroes fishing season: A decision is required on the period to which any TAC for the Faroes fishery would apply. The Faroes fishery has historically operated between October/November and May/June, but the historical TACs applied to a calendar year. This means that two different cohorts of salmon of each age class (e.g. two cohorts of 1SW salmon, etc.) were exploited under each TAC. ICES (2011) recommended that NASCO manage any fishery on the basis of fishing seasons operating from October to June, and catch advice should be provided on this basis.

Sharing agreement: The 'sharing agreement' establishes the proportion of any harvestable surplus within the NEAC area that could be made available to the Faroes fishery through the TAC. In effect, for any TAC option being evaluated for the Faroes, it is assumed that the total harvest would be the TAC divided by the Faroes share. WGNAS has proposed using a share allocation derived using the same approach and baseline period (1986–1990) as for West Greenland (ICES, 2010a). This gave a potential share allocation of 7.5% to Faroes. Following discussion within NASCO, one Party proposed an alternative baseline period of 1984–1988, which

would give a share allocation of 8.4% to Faroes. In the absence of further advice from NASCO, WGNAS has applied a value of 8.4%.

Choice of management units: ICES (2010a) noted that the stock complexes currently used for the provision of NEAC catch advice (southern NEAC and northern NEAC) are significantly larger than each of the six management units used for North American salmon (2SW only) in the catch advice for the West Greenland fishery. Basing an assessment of stock status on these large units greatly increases the risks to individual NEAC river stocks or groups of stocks that are already in a more depleted state than the average.

For the provision of catch advice on the West Greenland fishery, the total CL for NAC (2SW salmon only) of about 152 000 fish is assessed in six management units, which means that each unit has an average CL of about 25 000 salmon. In contrast, the total CLs for each of the NEAC stock complexes are:



The NEAC stock complexes are therefore between eight and 25 times the size of the average NAC ones. There is also wide variation in the size and status of stocks both within and among the NEAC national stock groups. WGNAS recommended (ICES, 2012a) that the NEAC catch advice should be based on more management units than are used at present, but also noted that there are practical limitations on the extent to which the assessments can be disaggregated, since the availability of information on the composition of the catch at Faroes constrained the selection of management units. In 2013, WGNAS (ICES, 2013) proposed a method to estimate the stock composition of the Faroes catch at a national level based on tag returns and the PFA estimates, but did not consider it appropriate to extend this to stock complexes smaller than this. Genetic stock assignment studies are underway to analyse scale samples collected at Faroes, but these are not expected to facilitate disaggregation below this level. In addition, other parameter values used in the assessment are currently only available for the total fishery and not smaller stock complexes.

In providing indicative catch advice with the new framework, WGNAS considered that it would be informative to managers to provide catch options tables for the ten NEAC countries as well as for the four stock complexes and has therefore run the risk framework using management units based on countries.

Management objectives: The management objectives provide the basis for determining the risks to stocks in each management unit associated with different catch options. The NASCO agreement on the adoption of a Precautionary Approach (NASCO, 1998) calls for the 'formulation of pre-agreed management actions in the form of procedures to be applied over a range of stock conditions', indicating that the management objectives (e.g. the required probability of exceeding the CL) should be agreed in advance of specific management proposals being considered.

At the request of NASCO, WGNAS considered the implications of applying probabilities of achieving CLs to separate management units vs. the use of simultaneous probabilities; this issue was outlined in detail in ICES (2013). The probability of simultaneous attainment of management objectives in a number of separate management units is roughly equal to the product of the probabilities of individual attainment for each management unit. The probability threshold for each individual management unit might reasonably be set at a fixed level unless there are specific reasons for adopting an alternative (e.g. for stock rebuilding). ICES (2012) recommended that an appropriate probability level for individual stock complexes would be 95%. This individual probability level can be applied to each management unit regardless of the number of units used; however, this is less obvious for the probability of simultaneous attainment, as explained next.

Management decisions for the West Greenland fishery have been based on a 75% probability of simultaneous attainment of CLs. For a given probability of achieving individual stock CLs, the probability of simultaneous attainment decreases rapidly as the number of management units considered increases. For the example of 20 management units (e.g. two age groups from each of ten countries), the use of the simultaneous probability level applied for West Greenland (75%) would correspond to the probability of individual stocks meeting the CLs being 98.6% or higher, assuming the same individual probability for all stocks. The use of a 95% probability level for meeting the CLs individually in the 20 management unit example, implies a simultaneous attainment probability of about 36%, i.e. there would be a 64% chance that at least one stock failed to meet its CL in any given year. On the other hand, the use of a 75% probability of simultaneous attainment could result in a fishery being advised when the individual probability of one management unit is as low as 75% if all the other management units have a 100% chance of meeting the CL (as in that case, the probability of simultaneous attainment would still be 75%). This may not be an acceptable risk for managing multiple river stocks.

WGNAS considered that the probability of simultaneous attainment can provide useful information to managers of the risk of failing to meet CLs in at least one stock in the MSF. However, as the management units being considered by NASCO for managing the MSF at Faroes are still very large and each unit encompasses a large number of individual river stocks, choosing a high probability level (such as 95%) of attaining CLs in individual units would be less risky to individual stocks than the use of a simultaneous attainment objective set at the value used for the West Greenland fishery.

On the basis of these considerations, WGNAS provided both individual probabilities and the probability of simultaneous attainment of the management units in the catch options tables (ICES, 2013). ICES recommends that management decisions should be based principally on a 95% probability of attainment of CLs in each stock complex individually. The simultaneous attainment probability may also be used as a guide, but managers should be aware that this will generally be quite low when large numbers of management units are used (as illustrated above, in the example with 20 management units).

3.4.3.2. Modelling approach for the catch options risk framework

The basic model for assessing each catch option within the risk framework is the same for both stock complexes and at a country level (ICES, 2013). The PFA forecasts derived in the Winbugs model are transferred to the risk framework model run in 'R'. The estimates and distributions of the PFA estimates used in the risk framework are derived by taking the first 50 000 values from the Winbugs posterior forecast simulations. Parameters in the following description that are marked with an '*' in the equations have uncertainty around them generated by means of 50 000 random draws

from the annual values observed from the sampling programmes conducted in the Faroes between the 1983/1984 and 1990/1991 fishing seasons. They therefore contribute to the estimation of the probability density function around the potential total harvest arising from each TAC option. When the assessment is run at a national level, the number of draws has to be limited to 25 000 because of memory limitations in 'R'.

The modelling procedure involves:

- estimating the total number of 1SW and MSW salmon that could be killed as a result of any TAC at Faroes, including catches in homewaters;
- adjusting these to their equivalent numbers at the time of recruitment to the Faroes fishery;
- subtracting these from the PFA estimates for maturing and non-maturing 1SW salmon in the appropriate years;
- assessing the results against the SERs (i.e. the CLs adjusted to the time of recruitment to the Faroes fishery).

The TAC option (T) is first divided by the mean weight (Wt*) of salmon caught in the Faroes fishery to give the number of fish that would be caught, and this value is converted to numbers of wild fish (Nw) by multiplying by one minus the proportion of fish-farm escapees in samples taken from the Faroes catch (pE*) observed in historical sampling programmes. A correction factor (C = 0.63) is applied to the proportion of fish-farm escapees to take account of reductions in the numbers of farm escapees over the past 20 years based on observations in Norwegian coastal waters:

 $Nw = T / Wt^* x (1 - (pE^* x C))$

This value is split into numbers by sea age classes (1SW and MSW) according to the proportion of each age group (pAi*, where 'i' is 1SW or MSW) observed in historical catch sampling programmes at Faroes. In the past, there has also been a requirement to discard any fish less than 60 cm total length caught in the Faroes fishery, and 80% of these fish were estimated to die, so these mortalities are also added to the 1SW catch. Thus:

Nw1SW = Nwtotal x pA1SW* + (Nwtotal x $pD^*/(1 - pD^*) \times 0.8)$

and

NwMSW = Nwtotal x pAMSW*

where: pD* is the proportion of the total catch that is discarded (i.e. fish <60 cm total length).

Further corrections are made to the 1SW and MSW numbers to reduce the 1SW total to take account of the proportion that will not mature as 1SW fish and to add the survivors from this group to the MSW fish in the following year. For the first catch advice year the number added to the MSW total is adjusted to the TAC of the current season (i.e. zero in 2012/2013). Thus:

Nw1SW = Nw1SW x pK *

and

 $NwMSW = NwMSW + Nw1SW \times (1-pK^*)$

where 'pK' is the proportion of 1SW salmon that are expected to mature in the same year (0.78) derived from experimental studies conducted in the 1980s (Youngson and Webb, 1993).

The numbers in each age group are then divided among the management units by multiplying by the appropriate proportions (pUij), where 'i' denotes the age groups and 'j' denotes the management units, and each of these values is raised by the Faroes share allocation (S) to give the total potential harvest (Hij) of fish from each management unit and sea age group:

 $Nwij = (Nwi \times pUij) / S$

Finally, these values are adjusted for natural mortality so that they can be compared with the PFA forecasts and SER values from the mid-date of the fishery to the recruitment date by using an instantaneous monthly rate of mortality of 0.03.

These harvests are then subtracted from the stock forecasts (PFAij) for the management units and sea age groups and compared with the Spawner Escapement Reserves (SERij) to evaluate attainment of the management objective. In practice, the attainment of the management objective is assessed by determining the probability that PFAij – Hij – SERij is greater than zero. The SER is the number of fish that need to be alive at the time of the Faroes fishery to meet the CL when the fish return to homewaters; this equals the CL raised by the mortality over the intervening time. CLs and SERs are currently estimated without uncertainty.

3.4.3.3. Input data for the risk framework

The analysis estimates probability of each management unit achieving its SER (the overall abundance objective) for different catch options in the Faroes fishery (from 0 to 200 t). The analysis assumes:

- no fishery operated in the 2012/2013 season;
- the TAC allocated to Faroes is the same in each year and is taken in full;
- homewater fisheries also take their catch allocation in full.

The analysis requires the following input data for the catch that would occur at the Faroes if a TAC was allocated (full details are provided in ICES, 2013):

- mean weights;
- proportion by sea age;
- discard rates (fish less than 60 cm total length);
- proportion of fish-farm escapees;
- composition of catches by management unit;
- proportion of 1SW fish not maturing.

3.4.3.4. Indicative catch advice

Table 3.4.3.4.1 provides an example of catch options for the Faroes fishery for the seasons 2013/2014 to 2015/2016 (ICES, 2013). Equivalent tables were provided for both 1SW and MSW salmon for all NEAC countries, and WGNAS also estimates the exploitation rates that these TAC options would impose on each stock complex or national stock (ICES, 2013). Updated catch options tables for the seasons 2015/2016 to 2017/2018 are provided in Section 3.5 of the 2015 Working Group report (see above).

Table 3.4.3.4.1. Probability of northern and southern NEAC - 1SW and MSW stock complexes
achieving their SERs independently and simultaneously for different catch options for the Faroes
fishery in the 2013/2014 to 2015/2016 fishing seasons.

Catch options	-	NEAC-N-	NEAC-N-	NEAC-S-	NEAC-S-	All
for 2013/14	(t)	1SW	MSW	1SW	MSW	complexes
season:	0	96.2%	99.8%	74.3%	75.6%	56.8%
	20	96.2%	99.2%	74.2%	69.8%	52.7%
	40	96.2%	98.2%	74.2%	63.9%	48.2%
	60	96.1%	96.3%	74.1%	57.9%	43.3%
	80	96.1%	93.4%	74.1%	52.1%	38.1%
	100	96.1%	89.3%	74.0%	46.6%	32.9%
	120	96.0%	84.3%	74.0%	41.7%	28.1%
	140	96.0%	78.4%	73.9%	36.8%	23.4%
	160	95.9%	71.6%	73.9%	32.5%	19.2%
	180	95.9%	64.6%	73.8%	28.5%	15.4%
	200	95.8%	57.6%	73.8%	25.0%	12.2%
Catch options	TAC option	NEAC-N-	NEAC-N-	NEAC-S-	NEAC-S-	All
for 2014/15	(t)	1SW	MSW	1SW	MSW	complexes
season:	0	94.6%	99.2%	75.4%	79.6%	59.0%
	20	94.6%	98.2%	75.3%	75.3%	55.8%
	40	94.6%	96.6%	75.3%	70.8%	52.0%
	60	94.5%	94.2%	75.2%	66.4%	48.0%
	80	94.4%	90.9%	75.2%	61.8%	43.6%
	100	94.4%	86.8%	75.1%	57.3%	38.9%
	120	94.3%	82.1%	75.1%	53.1%	34.4%
	140	94.3%	76.8%	75.0%	49.0%	30.1%
	160	94.3%	71.2%	75.0%	45.0%	25.9%
	180	94.2%	65.5%	74.9%	41.5%	22.1%
	200	94.2%	59.6%	74.9%	38.0%	18.6%
Catch options	TAC option	NEAC-N-	NEAC-N-	NEAC-S-	NEAC-S-	All
for 2015/16	(t)	1SW	MSW	1SW	MSW	complexes
season:	0	94.6%	98.5%	70.1%	79.7%	55.2%
	20	94.6%	97.2%	70.1%	76.0%	52.4%
					FO 0 0/	49.2%
	40	94.5%	95.1%	70.0%	72.2%	17.270
X	40 60		95.1% 92.3%	70.0% 70.0%	72.2% 68.4%	45.6%
X		94.5% 94.5% 94.5%				
	60 80	94.5% 94.5%	92.3% 89.0%	70.0% 69.9%	68.4% 64.6%	45.6% 41.9%
	60	94.5%	92.3%	70.0%	68.4%	45.6%
	60 80 100 120	94.5% 94.5% 94.4% 94.4%	92.3% 89.0% 85.0% 80.6%	70.0% 69.9% 69.9% 69.8%	68.4% 64.6% 60.7% 57.1%	45.6% 41.9% 38.0% 34.2%
	60 80 100 120 140	94.5% 94.5% 94.4% 94.4% 94.3%	92.3% 89.0% 85.0% 80.6% 75.7%	70.0% 69.9% 69.8% 69.8%	68.4% 64.6% 60.7% 57.1% 53.5%	45.6% 41.9% 38.0% 34.2% 30.4%
	60 80 100 120	94.5% 94.5% 94.4% 94.4%	92.3% 89.0% 85.0% 80.6%	70.0% 69.9% 69.9% 69.8%	68.4% 64.6% 60.7% 57.1%	45.6% 41.9% 38.0% 34.2%

3.5. Development of indicator frameworks to identify significant changes in previously provided multiannual management advice

3.5.1. Background

In support of the multiannual management advice that is provided for all three NASCO Commission Areas, NASCO asked ICES to provide an assessment of the

minimal information needed to signal an unforeseen change in productivity for stocks contributing to fisheries within each Commission area. A particular concern was that an increase in productivity may alter the reliability of the previously provided multiyear catch options and could result in unrealised harvest within various mixed-stock fisheries. Initial progress on this issue was presented to WGNAS in 2006 (ICES, 2006) and further developments were made by the Study Group on Establishing a Framework of Indicators of Salmon Stock Abundance [SGEFISSA] which met in November 2006 (ICES, 2007b) and reported to WGNAS in 2007 (ICES, 2007a). This resulted in the development of a suggested framework (Framework of Indicators - FWI) which could be used to indicate if any significant change in the status of stocks had occurred and thus confirm whether the previously provided multi-annual management advice was still appropriate.

The initial FWI was developed with both the Greenland and Faroes fisheries in mind, although the methodology only proved suitable for the West Greenland fishery and an alternative approach was subsequently developed for the NEAC area (ICES, 2011; 2012a; 2013). Thus, FWIs are now routinely applied in the interim (non-assessment) years of multiyear agreements for both NAC and NEAC to facilitate the management of the West Greenland and Faroes fisheries respectively. Both operate according to the timeline outlined in Figure 3.5.1.1. Outline descriptions of the two different schemes are provided below.

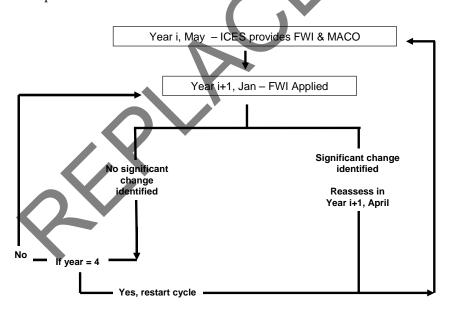


Figure 3.5.1.1. Timeline for employment of the Framework of Indicators (FWI). In Year i, ICES provides an updated FWI which re-evaluates the updated datasets and is summarized in an Excel worksheet. In January of Year i+1 the FWI is applied and two options are available depending on the results. If no significant change is detected, no re-assessment is necessary and the cycle continues to Year i+2. If no significant change is detected in Year i+2, the cycle continues to Year i+3. If a significant change is detected in any year, then reassessment is recommended. In that case, ICES would provide an updated FWI the following May. ICES would also provide an updated FWI if year equals 4. [MACO = multi-annual catch options].

3.5.2. Framework of Indicators (FWI) for the West Greenland Fishery

The process for developing and applying the FWI for the Greenland fishery consists of six general steps:

- <u>Definition of a significant change</u> Define measurable criteria for what the statement "a significant change in the previously provided multi-annual management advice" represents.
- <u>Evaluating historical relationships between indicators and variables of in-</u> <u>terest</u> - Define and evaluate the historical relationships between numerous indicators and the variable of interest for individual rivers across all stock complexes.
- <u>Establishing threshold values</u> Define the threshold level (i.e. variable of interest level) that will satisfy the management objectives for each stock complex.
- <u>Decision rule determinations</u> Define and apply a standardised approach for determining the appropriate decision rule value. The decision rule should provide a signal if the variable of interest will be greater than or less than the threshold level with high precision.
- <u>Combining Indicators within the Framework</u> Define and apply a standardised approach for combining indicator datasets within and across stock complexes for future comparison against contemporary indicator values.
- <u>Applying the FWI</u> Define and apply a standard approach to input contemporary indicator values into the FWI to determine if there is likely to be a significant change in the previously provided management advice.

Each of these is considered in turn; full details are available in ICES (2007b).

3.5.2.1. Definition of a significant change

A significant change in the previously provided multiannual management advice is regarded as an unforeseen change in stock status that would alter the previously provided advice based on analysis of current population data obtained from various monitored populations across the North Atlantic. This would be indicated by a situation where stock abundance has increased to a level where a fishery could be recommended when no catch had previously been advised, or a decrease in stock abundance when catch options had been chosen.

For the fishery at West Greenland, ICES would recommend that a harvestable surplus exists within the West Greenland stock complex if there was a high probability (75%) that the following three objectives could be met simultaneously:

The conservation limits of the four northern regions of North America (Labrador, Newfoundland, Québec, and Gulf) were achieved.

There was a 25% increase in returns to the Scotia-Fundy region relative to the mean returns for the 1992–1996 period. For the USA, the management objective was revised in 2014 to correspond to recover objectives defined in the recovery plan for endangered Atlantic salmon stocks in the USA (ICES, 2014), this now requires that estimated 2SW adult returns are 4549 or greater.

The conservation limit for the Southern NEAC MSW complex was achieved.

3.5.2.2. Evaluating historical relationships between indicators and variable of interest

A number of variables were considered for inclusion as indicators in the FWI, but only two were considered sufficiently informative to be carried forward into the framework: adult returns (returns, catch or estimated PFA) and return rates (i.e. smolt survival rates, marine survival). These are available, by sea age class, for a number of monitored rivers throughout the North Atlantic and can be directly related to the management objectives for the fishery.

3.5.2.3. Establishing threshold values

In keeping with the 75% probability of meeting or exceeding the objectives for the West Greenland catch options (see above), the 25th percentile of the return estimates of the six areas in North America are compared to the corresponding 2SW conservation limits of the four northern areas of North America, to the 25% increase objective for the Scotia- Fundy area, and to management objective of achieving 4549 or greater 2SW adult returns for the USA. For the southern NEAC non-maturing component, the 25th percentile of the PFA estimate of the southern NEAC non-maturing complex is compared to the spawning escapement reserve (SER) for the southern NEAC non-maturing complex.

3.5.2.4. Decision rule determinations

The procedure for analysing the relationships between the indicators and the returns of 2SW salmon or the PFA estimates was originally suggested by ICES (2006). The individual river catches, returns or return rates are lagged to correspond to the same smolt cohort for the 2SW returns to North America or to the PFA estimates for NEAC complexes. Bivariate plots of each indicator dataset relative to the 2SW returns or the PFA estimates are prepared. Upper and lower halves are defined by the management objective value for the corresponding geographic area in North America or the NEAC stock complexes as outlined above. Estimates of returns of 2SW or PFA estimates in the upper half correspond to years when the returns or PFA exceed the management objectives. Points in the lower half correspond to years when the returns or PFA are less than the management objective.

Left and right halves are defined by a sliding rule along the indicator range. An objective function that maximises the number of correct assignments (true highs and true lows) is used to define the indicator decision rule. The objective function also minimises the number of incorrect assignments (false highs and false lows, Figure 3.5.2.4.1).

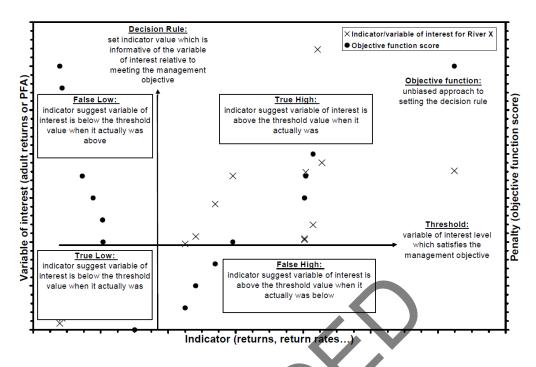


Figure 3.5.2.4.1. Example of Indicator/Variable of Interest exploratory graph identifying the threshold value, decision rule, penalty function and the four states (true high, true low, false high and false low).

The value of the indicator variable that minimises the sum of the penalty scores (i.e. minimises the number of incorrect assignments) is assigned as the decision rule for that dataset. Equal penalty weights are assigned to false highs (lower right quadrant) and false lows (upper left quadrant). Correct assignments are scored as zero. In the case when multiple minima occurred, the lowest indicator value among the low minima values is chosen.

Indicators are retained in the framework when they are evaluated as being informative of the magnitude of returns or PFA relative to the management objectives. These informative indicator datasets also have to meet the following two criteria to be retained:

- Expectation that the indicator variable would be available in future (in January), and
- A minimum of five observations are present in each of the correct quadrats (true low; true high).

3.5.2.5. Combining Indicators within the Framework

The probabilities of correct assignments are calculated for each of the true low and true high states for each of the indicator datasets retained. The respective probabilities correspond to the ratio of the correct assignment to all observations within the respective low and high indicator halves:

P(Statelow | Indicatorlow) (i.e. true low) = N(Statelow | Indicatorlow) / N Indicatorlow

P(Statehigh | Indicatorhigh) (i.e. true high) = N(Statehigh | Indicatorhigh) / N Indicatorhigh

Indicator datasets are then pooled according to management objective/stock complex groupings. Each NAC stock complex (n=6) and the southern NEC non-maturing stock complex are pooled separately as these stock complexes relate to the management objectives for the West Greenland fishery.

3.5.2.6. Applying the FWI

To apply the FWI, the most recent year's indicator value for each of the retained indicator datasets is compared to the decision rule as determined from the historical datasets. If the contemporary indicator value is low relative to the decision rule, it is assigned a value of -1. If the value is high, it is assigned a score of +1. Multiple indicators within the stock complex groupings are then combined by arithmetic average of the product of the indicator value (-1, +1) and the probability of a correct assignment corresponding to the true low or true high states. An average geographic area or stock complex score equal to or greater than zero suggests there is a likelihood of meeting the management objective for that grouping based on the historic relation between the variable of interest (adults returns or PFA) and the indicators evaluated.

If the scores for all the groupings within a fishery complex are greater than zero, then there is a likelihood that all the management objectives for that fishery will be met. Under that scenario, the multiyear management advice should be reassessed. When the score(s) for one or more of the groupings is less than zero, there is unlikely to be a significant change in the management advice and there would be no need for a reassessment.

SGEFISSA (ICES, 2007b) developed a spreadsheet template FWI (see example at Figure 3.5.2.6.1) in which the underlying variable of interest/ indicator dataset relationships and decision rules are summarised and collated according to the specific management objectives for each fishery. This provides one of two conclusions for the user:

- 1) No significant change identified by the indicators;
- 2) Reassess.

	Overall Recommendation No Significant Change Identified by Indicators										
		2011	Ratio Value to				Indicator	Probability of Correct	Indicator	Managemer Objective	
Beographic Area	River/ Indicator	Value		Threshold			State	Assignment	Score	Met?	
ISA	Penobscot 2SW Returns Penobscot 1SW Returns	2368 741	167% 197%	1415 377	100% 83%	92% 88%	1 1	0.92 0.88	0.92 0.88		
	Penobscot 2SW Survival (%)	0.39	170%	0.23	100%	60%	1	0.6	0.6		
	Penobscot 1SW Survival (%)	0.12	133%	0.09	85%	73%	1	0.73	0.73		
	Narraguagus Returns	196	196%	100	95%	61%	1	0.61	0.61		
	possible range				-0.93	0.75					
	Average		173%						0.75	Yes	
		00.4	00/	0.000	000/	4000/			0.00		
cotia-Fundy	Saint John Return Large	294 146	9% 51%	3 329 285	96% 77%	100% 85%	-1	0.96 0.77	-0.96 -0.77		
	Lahave Return Large St. Mary's Return Large	146	6%	285	100%	85% 73%	-1 -1	0.77	-0.77		
	North Return Large	1 193	168%	712	95%	67%	-1	0.67	0.67		
	Saint John Return 1SW	582	26%	2 276	86%	80%	-1	0.86	-0.86		
	LaHave Return 1SW	565	34%	1 679	94%	67%	-1	0.94	-0.94		
	St. Mary's Return 1SW	331	16%	2 038	95%	93%	-1	0.95	-0.95		
	Saint John Survival 2SW (%)	0.13	59%	0.22	95%	81%	-1	0.95	-0.95		
	Lahave Survival 2SW (%)	0.88	367%	0.24	81%	81%	1	0.81	0.81		
	Saint John Survival 1SW (%)	0.12	16%	0.76	86%	73%	-1	0.86	-0.86		
	Lahave Survival 1SW (%)	0.72	50%	1.44	92%	78%	-1	0.92	-0.92		
	Liscomb Survival 2SW (%)	0.03	60%	0.05	86%	91%	-1	0.86	-0.86		
	East Sheet Harbour Survival 2SW (%)	0.005	25%	0.02	67%	82%	-1	0.67	-0.67		
	possible range Average		68%		-0.88	0.81			-0.64	No	
	Average		0070						-0.04	No	
	Miramichi Return 2SW	28 977	183%	15 800	100%	85%		0.85	0.85		
	Miramichi Return 1SW	45 880	110%	41 790	89%	67%	1	0.67	0.67		
	possible range				-0.95	0.76					
	Average		147%						0.76	Yes	
luebec	Cascapédia Return Large	3 815	167%	2 280	69%	92%		0.92	0.92		
luebec	Bonaventure Return Large	1 259	85%	2 280	69% 75%	92% 81%		0.92	-0.75		
	Grande Rivière Return Large	533	121%	442	100%	94%	. 1	0.94	0.94		
	Saint-Jean Return Large	688	91%	758	86%	89%	A 1	0.86	-0.86		
	Dartmouth Return Large	1 171	155%	756	86%	89%	1	0.89	0.89		
	Madeleine Return Large	996	153%	653	70%	93%	1	0.93	0.93		
	Sainte-Anne Return Large	871	201%	433	67%	88%	1	0.88	0.88		
	Godbout Return Large	694	108%	641	86%	100%	1	1	1		
	De la Trinite Return Large	317	82%	385	75%	100%	-1	0.75	-0.75		
	York Return Return Large	1 585	113%	1405	63%	83%	1	0.83	0.83		
	Grande Rivière Return Small	237		199	59%	80%	1	0.8	0.8		
	Saint-Jean Return Small	343		394	53%	80%	-1 1 1	0.53	-0.53		
	Godbout Return Small	623	123%	508	85%	92%		0.92	0.92		
	De la Trinite Return Small	949	238%	399		89% 83%		0.83	0.83		
	De la Trinite Survival Large (%)	0.76 2.54	155% 170%	0.49 1.49	88% 63%	96% 89%	1 1	0.96 0.89	0.96 0.89		
	De la Trinite Survival Small (%) Saint-Jean Survival Small (%)	1.86	258%	0.72	100%	64%	1	0.69	0.69		
	possible range	1.00	23070	0.72	-0.77	0.478		0.04	0.04		
	Average		143%						0.50	Yes	
lewfoundland	Exploits Return Small	34 085	137%	24 924	83%	56%	1	0.56	0.56		
	Middle Brook Return Small	2 642	141%	1 868	84%	63%	1	0.63	0.63		
	Torrent Return Small	2 784	67%	4 154	94%	64%	-1	0.94	-0.94		
	possible range Average		115%		-0.87	0.61			0.08	Yes	
	Average		115%						0.08	162	
abrador											
	possible range										
	Average								NA	Unknow	
outhern NEAC	possible range										
	DUSSIDIE Idilue										

Figure 3.5.2.6.1. Framework of indicators spreadsheet for the West Greenland fishery. For illustrative purposes, the 2011 value of returns or survival rates for the 40 retained indicators is entered in the cells corresponding to the annual indicator variable values.

If no significant change has been identified by the indicators, then the multiyear catch advice for the year of interest could be retained. If a significant change is signalled by the indicators, the response is to reassess.

The framework spreadsheet is designed to capture both fishing and non-fishing scenarios:

- Multiyear advice provides no catch options greater than zero but indicators are suggesting that the management objectives may be met (conclusion: Reassess);
- Multiyear advice provides catch options greater than zero but the indicators suggest the management objectives may not be met (conclusion: Reassess).

There are two steps required by the user to run the framework. The first step in the framework evaluation is to enter the catch advice option (i.e. tonnes of catch) for the West Greenland fishery. This feature provides the two way evaluation of whether a change in management advice may be expected and a reassessment would be required. The second step is to enter the values for the indicator variables in the framework for the year of interest. The spreadsheet evaluation update is automated and the conclusion is shown in the row underneath "Overall Recommendation".

The conclusions from the framework evaluation are based on whether there is simultaneous achievement of the management objectives in the six stock areas of North America and the southern NEAC 1SW non-maturing complex. If there are no indicator variables for a geographic area, the attainment of the management objectives is evaluated as unknown and that area or complex is not used in the decision structure of the framework.

Within the geographic areas for which indicator variables are retained, all the available indicators are used to assess the indicator score. If an update value for an indicator variable is not available for the year of interest, the indicator variable is not used to quantify the indicator score for that area.

The West Greenland FWI was updated during the 2015 WGNAS meeting and an updated spreadsheet produced. Details are provided in Section 5.8 of the Working Group report (see above).

3.5.3. Framework of Indicators (FWI) for the Faroes Fishery

3.5.3.1. Background

The original FWI applied to the West Greenland fishery (ICES, 2007b) was not applicable for the Faroese fishery for a number of different reasons. Among these were the lack of quantitative catch advice, the absence of specific management objectives and a sharing agreement for this fishery and the fact that none of the available indicator datasets met the criteria for inclusion in the FWI.

In 2011, WGNAS re-evaluated the approach for developing a FWI for the Faroese fishery (ICES, 2011). Since over the available time-series the PFA estimates for the NEAC stock complexes have predominately remained above the SER, the Working Group suggested a different set of decision rules for this FWI. It was suggested that the status of stocks should be re-evaluated if the FWI suggests that the PFA estimates are deviating substantially from the median values from the forecast.

3.5.3.2. Description of the FWI

It was initially suggested that the 95% confidence interval range for the mean of the indicator prediction, relative to the median forecast value, be used to compute the decision thresholds for whether the indicator suggests a reassessment or not (ICES, 2011). The limits should be computed at the median values of the PFA forecasts in each of the years in multiyear advice. However, the 95% criterion was subsequently re-examined (ICES, 2012a) and it was recommended that the upper and lower 75% confidence limits of the individual predictions be used for comparison (Figure 3.5.3.2.1). WGNAS recognised that this was a relaxation of the decision rule suggested in 2011, and will lead to a larger interval, and thus a lower chance of a reassessment than the approach suggested in 2011. However, this was considered to be a more realistic confidence level given the relatively wide variability of the indicator datasets, and was also consistent with the approach adopted by NAC.

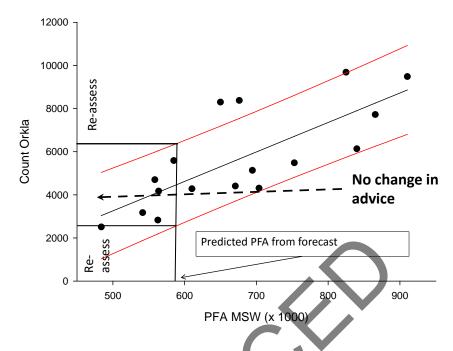


Figure 3.5.3.2.1. Example of how the reassessment intervals for the indicators are computed. The values of an indicator (counts) are plotted against the PFA. The regression line is shown in black and 75% confidence limits for the individual estimates are shown in red. From the forecasted PFA in the year in question the values of the indicator corresponding to the upper and lower 75% confidence interval are estimated. If the indicator value falls outside these limits a reassessment is recommended by this particular indicator.

When the stocks are divided into smaller management units, potential indicators for each management unit become relatively scarce. Therefore, the Working Group recommended that the FWI be regressed against the stock complexes that they belong to. For example, MSW indicators from Norway should be regressed against PFA MSW for northern NEAC.

In 2012, the FWI was applied as a two-tailed test (ICES, 2012a). However, it was subsequently agreed that, in the event of a closed fishery, the indicators should only be compared to the upper 75% confidence limit (i.e. a one-tailed test). This means that for a closed fishery, a reassessment is only triggered where the forecast appeared to be an underestimate and there may be a possibility of a harvest being denied. In the case of an open fishery they should be compared to both the upper and lower 75% confidence limits. In this case, if the FWIs suggest that the forecasted PFA is either an underestimation or an overestimation of the realised PFA in any of the four stock complexes, then this should trigger a reassessment.

ICES further advised (ICES, 2015) that, in the case of closed fisheries, the FWI should only be applied to those stock complexes which had previously signalled zero catch options at Faroes. This was agreed by NASCO (NEA(16)11) and applied in January 2017.

WGNAS developed a FWI spreadsheet (ICES, 2011) to provide an automatic evaluation of the need for a reassessment once the new indicator values are available in January; this has been updated in subsequent years (ICES, 2012a; 2013, 2015). An example spreadsheet is provided at Figure 3.5.3.2.2. The following summarizes the main steps performed by the spreadsheet following updating of the relevant data for the variable of interest by adding the latest year's number:

- Regression analysis with the dataset *x* to determine its power to predict PFA in the forecasted years.
- Calculation of the 75% confidence intervals of individual predictions of the regression for dataset *x*. An indicator value below the 75% individual confidence interval (CI) is interpreted as indicative of an overestimation of the PFA, while a point above the 75% individual confidence interval is interpreted as indicative of an underestimation of PFA.
- A dataset is considered informative and should be kept as an indicator in the FWI if the following conditions are met: sample size $(n) \ge 10$; $r^2 \ge 0.2$; dataset updated annually and new value available by January 15. Datasets that do not meet these criteria are discarded.
- Apply a binary score to each indicator value. Thus, for dataset *x*, if the current year's indicator value is outside the 75% individual regression point estimate CI (below or above) then that indicator receives a score of 1. If the indicator is within the 75% CI, then the indicator receives a score of -1. In the absence of an indicator datapoint for any year, a score of zero is applied. Whether the indicator value is above or below the upper and lower CI values is checked separately in two spreadsheet columns and a decision whether the indicator value is within the *CI* is assessed by combining the information in the two columns.
- Separate columns are used to sum the scores for all the indicator datasets within each stock complex. This is done separately for points that fall above the CI and those that fall below. In the case of a two-sided approach (open fishery), if the sum of these columns is ≥0, then the spreadsheet signals "REASSESS"; if the sum is <0, then it signals "No significant OVERestimation of PFA identified by indicators, do not reassess" for indicator values that fall below the CI, and "No significant UNDERestimation of PFA identified by indicators, do not reassess" for indicator values that are above the CI. In case of a one-sided approach (closed fishery), only underestimation will signal a "REASSESS".
- FWI results are generated for each stock complex (northern NEAC maturing and non-maturing, and southern NEAC maturing and non-maturing).
 A score of ≥0 for any of these stock complexes would signal a reassessment.

WGNAS reassessed the effects of applying stricter criteria than $r^2 \ge 0.2$ for inclusion of indicators in the FWI. As stricter criteria are used, the number of indicators included reduces rapidly. It was therefore concluded to keep the criterion of $r^2 \ge 0.2$ in order to obtain a sufficient number of indicators to be able to use the FWI even in the event of one or more indicators being unavailable by the time the FWI is applied each year. The r^2 value of 0.2 corresponds to a value slightly lower than what is considered to be a "large" effect size (r = 0.5, $r^2 = 0.25$) by Cohen (1988). Although a criterion of $r^2 \ge 0.2$ gives each indicator little predictive power alone (Prairie, 1996), the approach of using a suite of indicators is more similar to meta-analysis (Rosenthal, 1984) meaning that the outcome of the FWI is not dependent on the result of one indicator in isolation, but rather on the combined performance of the indicator set.

The Faroes FWI was updated during the 2015 WGNAS meeting and an updated spreadsheet produced. Details are provided in Section 3.7 of the Working Group report (see above). In evaluating all the time-series, it was noted that the lower 12.5 % CL, which is used to determine which indicator values are outside the 75% CI on the lower side, was negative for some regression relationships for predicted PFA values in 2015 and 2016. Since this would invalidate the use of such indicators (they would not indicate that predicted PFA values are overestimates regardless of how small they are), an additional (fourth) criterion was established as a requirement for including time-series in the FWI. This requires that the lower 12.5% confidence limit for an indicator time-series should be positive for any values of PFA included in the FWI.

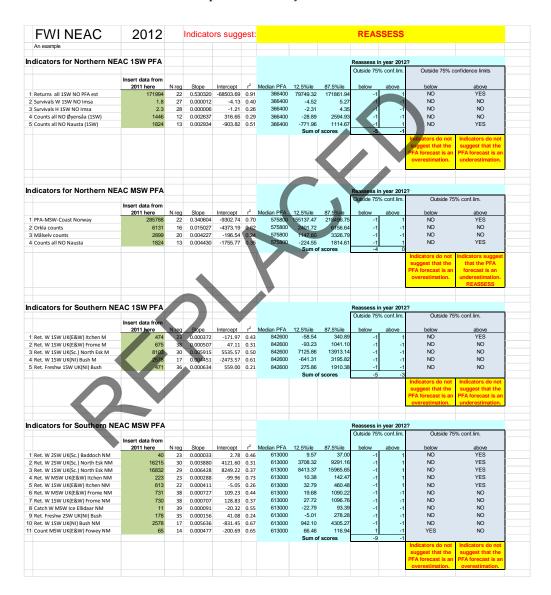


Figure 3.5.3.2.2. Output of the spreadsheet for the test of FWIs for NEAC for 2012 based on the values of the indicators from 2011. Because the indicators suggest that the forecast for northern NEAC MSW was an underestimate, the overall advice from the spreadsheet is reassess.

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Appendices to Stock Annex

Appendix 1 (a): Description of how catch-and-release mortality is incorporated in regional and national stock assessments

Commission			
Area	Country/Region	How it is used in regional and national assessments	Future developments / improvements
NAC	Canada-Quebec	C&R has become more popular in the region and C&R only angling licenses are sold. C&R data are incomplete as there is no requirement to report C&R numbers. Generally, C&R mortality is considered in the assessment but the majority of the assessments are conducted as spawner counts after the fisheries so any losses due to C&R mortality are accounted for in the spawner estimates but not in the returns (which are the sum of known losses and spawning escapement).	New studies of the contribution of C&R fish to spawning success have been initiated. C&R monitoring is becoming more complete. Consideration will be given in the future to incorporating these losses in the returns and in the assessments based on angling catches, especially as reporting improves.
	Canada- Newfoundland & Labrador	Catch and release mortality is included in estimates of spawners. Spawning escapement is reduced by 5-15% (mean 10%) of the released catch.	No plans for further development.
	Canada - Gulf	Assessments of spawners are adjusted by mortality rates of 3% to 6% of the total C&R estimates of small and large salmon. The rates vary by river according to angling seasons, and the occurence of other factors such as disease which can affect survival of salmon.	Catch and release mortality is known to be affected by the water temperatures when fish are angled. In some cases, angling fisheries are closed when water temperatures are high in the summer to reduce the losses of fish from C&R. Methods to determine catch and release numbers vary by river and in some cases, the number of released fish is estimated from returns and historical creel survey data. As the practice of C&R becomes more popular, estimation methods for C&R values will have to be revisited.
	Canada – Scotia/Fundy	Assessments are currently adjusted by 4% of the C&R fish to correct for C&R mortality.	Numbers of C&R fish are currently low (retention fisheries are closed). If C&R catches increase, further research on the correction factor would be warranted.
	USA	No correction for mortality due to C&R used in estimating spawner numbers. However, all fisheries have been closed and the number of fish caught relative to stock size is very small.	
NEAC	Russia	With increasing C&R the retained catch for similar	
	Norway Sweden	effort is reduced. Therefore the exploitation rate for retained fish is lower. The increase in C&R in recent years is incorporated into the national run-	If C&R information is incorporated into formal assessments then muliple recaptures should be taken into account. C&R
	Iceland UK(Scotland)	reconstruction model by reducing the exploitation rate value used in the model input. This is assessed qualitatively. No correction for increased C&R mortality	mortality should be incorporated into estimates of spawning escapement.
	Ireland	is applied when estimating the spawning escapement. No correction for mortality due to C&R used in estimating spawner numbers or in the national run- reconstruction model.	Incorporation of formal method for estimating the effect of C&R on number of returning fish. Incorporation of C&R mortality in estimates of spawning escapement
	UK(England & Wales)	With increasing C&R the retained catch for similar effort is reduced. Therefore the exploitation rate for retained fish is lower. The increase in C&R in recent years is incorporated into the national run- reconstruction model by reducing the exploitation rate value used in the model input. This is assessed qualitatively. 20% mortality of C&R fish used in assessing compliance with river-specific conservation	If C&R information is incorporated into formal assessments then multiple recaptures should be taken into account.
	UK(N. Ireland)	Returns are estimated by raising the reported net catch by exploitation rate. No correction for increased C&R mortality is applied when estimating the spawning escapement.	If C&R information is incorporated into formal assessments then multiple recaptures should be taken into account. C&R mortality should be incorporated into estimates of spawning escapement.
	Denmark	C&R rates recorded, but no national run-reconstruction assessment applied.	
	Finland	No record of C&R	If C&R information is collected, it should be incorporated into formal assessments and multiple recaptures should be

Appendix 1 (b): Description of how unreported catch is incorporated in
regional, national and international stock assessments

Canada-Quebec Canada- Vewfoundland & Jabrador Canada - Gulf Canada - Scotia/Fundy	How it is used in regional and national assessments Unreported catches are based on historical estimates relative to stock size or are provided by field conservation and protection staff. Unreported catches when available are included in the regional assessment of returns and spawners. Catch statistics include estimates of harvests by log book non-respondents. Therefore they are included in the regional assessments and the PFA estimate. No account is taken of illegal fisheries. Unreported catches are sometimes provided by Conservation and Protection Personnel and are estimates of illegal fishing removals within specific regions. Unreported catches have not been used in the assessments of returns or spawners. No adjustment made, with the exception of the Saint John River where returns/spawners are adjusted for estimated bycatch and poaching. In other rivers where assessments directly quantify spawners, returns would be underestimated if fatch is under reported. Unreported catch is estimated to be zero and therefore has no effect on national assessments.	Unreported catches which occur in marine waters outside the jurisdiction of the regions are not included in the run reconstruction models.	Future developments / improvements If unreported catch estimates were provided they could be incorporated in the regional assessments and in the continent estimates of abundance and spawners. Unreported catch may be accounted for in either the returns or the spawners, depending upon when and where the illegal activity occurs relative to the location and time of the assessment model.
Canada- Newfoundland & .abrador Canada - Gulf Canada – icotia/Fundy JSA Russia	relative to stock size or are provided by field conservation and protection staff. Unreported catches when available are included in the regional assessment of returns and spawners. Catch statistics include estimates of harvests by log book non-respondents. Therefore they are included in the regional assessments and the PFA estimate. No account is taken of illegal fisheries. Unreported catches are sometimes provided by Conservation and Protection Personnel and are estimates of illegal fishing removals within specific regions. Unreported catches have not been used in the assessments of returns or spawners. No adjustment made, with the exception of the Saint John River where returns/spawners are adjusted for estimated bycatch and poaching. In other rivers where assessments directly quantify spawners, returns would be underestimated if catch is under reported. Unreported catch is estimated to be zero and therefore	Unreported catches which occur in marine waters outside the jurisdiction of the regions are not included in the run reconstruction models.	provided they could be incorporated in the regional assessments and in the continent estimates of abundance and spawners. Unreported catch may be accounted for in either the returns or the spawners, depending upon when and where the illegal activity occurs relative to the location and time of the
Newfoundland & abrador Canada - Gulf Canada - Canada - Canada - Canada - Canada - Canada - Sanada - Sa	book non-respondents. Therefore they are included in the regional assessments and the PFA estimate. No account is taken of illegal fisheries. Unreported catches are sometimes provided by Conservation and Protection Personnel and are estimates of illegal fishing removals within specific regions. Unreported catches have not been used in the assessments of returns or spawners. No adjustment made, with the exception of the Saint John River where returns/spawners are adjusted for estimated bycatch and poaching. In other rivers where assessments directly quantify spawners, returns would be underestimated if catch is under reported. Unreported catch is estimated to be zero and therefore	ed on historical estimates provided by field is taff. Unreported catches lin the regional assessment nates of harvests by log refore they are included in d the PFA estimate. No iheries. etimes provided by Personnel and are envoals within specific is have not been used in the nawners. he exception of the Saint pawners, returns would s under reported. et to be zero and therefore isessments. imates of the unreporting ional PFA estimates from mon. National estimates (which incorporate in spawners), returns would s under reported. Assumed to be negligible unreported catches spawner stimates for stock complexes. Assumed to be negligible unreported catches sime of in stock assessments. Annual va estimates for both the NAC and NEAC areas. Since 1993, this has been provided by the estimates for both the NAC and NEAC areas.	provided they could be incorporated in the regional assessments and in the continent estimates of abundance and spawners. Unreported catch may be accounted for in either the returns or the spawners, depending upon when and where the illegal activity occurs relative to the location and time of the
Canada – icotia/Fundy JSA Russia	Conservation and Protection Personnel and are estimates of illegal fishing removals within specific regions. Unreported catches have not been used in the assessments of returns or spawners. No adjustment made, with the exception of the Saint John River where returns/spawners are adjusted for estimated bycatch and poaching. In other rivers where assessments directly quantify spawners, returns would be underestimated if catch is under reported. Unreported catch is estimated to be zero and therefore	marine waters outside the jurisdiction of the regions are not included in the run reconstruction models.	spawners. Unreported catch may be accounted for in either the returns or the spawners, depending upon when and where the illegal activity occurs relative to the location and time of the
Scotia/Fundy JSA Russia	John River where returns/spawners are adjusted for estimated bycatch and poaching. In other rivers where assessments directly quantify spawners, returns would be underestimated if catch is under reported. Unreported catch is estimated to be zero and therefore		
Russia			
Norway			
Sweden		National estimates (which incorporate	Incorporate revised estimates of
	Minimum and maximum estimates of the unreporting		minimum and maximum estimates of
			unreporting rate as national estimates
	the catch of 1SW & MSW salmon.		are improved.
	No national assessment		
		Assumed to be negligible unreported	Sampling programme if fishery
		1SW fish is incorporated in stock	resumes.
N. Greenland	Not applicable	Unreported catch at West Greenland is incorporated in assessments for both the NAC and NEAC areas. Since 1993, this has been provided by the	Annual variation in unreported catch estimates would be incorporated into the model.
0			
	veden eland ((Scotland) ((England & ales) ((N. Ireland) ance enmark roes	veden Minimum and maximum estimates of the unreporting eland rate are used in deriving national PFA estimates from (Scotland) the catch of 1SW & MSW salmon. ((England & ales)) ales) ance ance enmark No national assessment roes Not applicable	veden National estimates (which incorporate unreported ctaches) are aggregated to provide PFA, return and spawner estimates for btb unreported ctaches) are aggregated to provide PFA, return and spawner estimates for stock complexes. (Scottand) the catch of 1SW & MSW salmon. (Ingland & ales) (International PFA estimates from on the catch of 1SW & MSW salmon. (KI. Ireland) ance enmark No national assessment roes Not applicable Assumed to be negligible unreported in stock assessments. . Greenland Unreported ctach at West Greenland is incorporated in assessments for both the NAC and NEAC areas. Since 1993, this has been provided by the Greenlandic authorities. Prior to this time, no unreported catch component

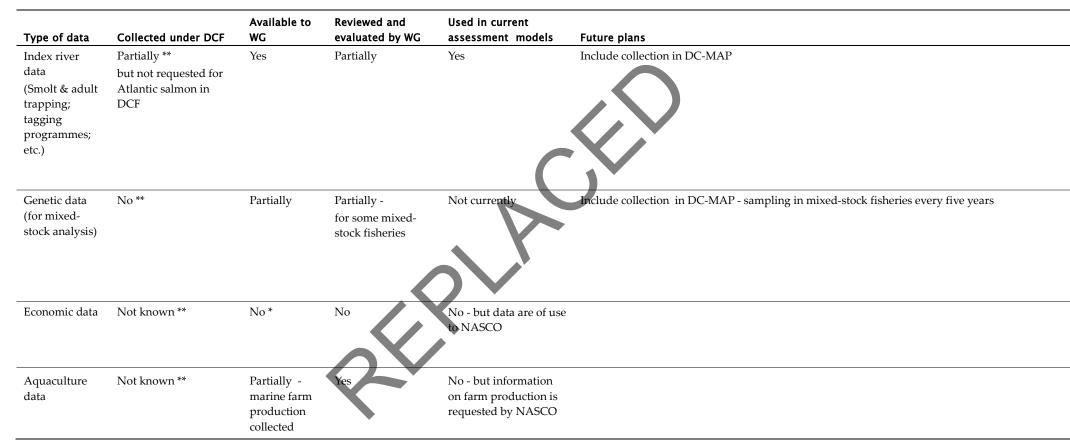
Appendix 2: Overview of current DCF and future data needs for Atlantic salmon assessment/ scientific advice

Type of data	Collected under DCF	Available to WG	Reviewed and evaluated by WG	Used in current assessment models	Future plans
How to be	Yes/	Yes/	Yes/	Yes/	Keep as current DCF/ Improve sampling intensity/ No need to be collected/ (other free text)
filled	No/	No/	No/	No/	
	Partially	Partially	Partially	Partially used	
Fleet capacity	No **	No *	No	No	No need to be collected
Fuel consumption	No **	No *	No	No	No need to be collected
Fishing gear and effort	Partially **	Partially	Partially	Partially, but information requested by NASCO	Use for estimation of exploitation rates. Improve coverage and sampling intensity in DC-MAP
Landings	Partially **	Yes	Yes	Yes	Improve coverage in DC-MAP

Type of data	Collected under DCF	Available to WG	Reviewed and evaluated by WG	Used in current assessment models	Future plans
Discards	No **	No *	No	No	No need to be collected
Recreational fisheries	Partially **	Yes	Yes	Yes	Improve coverage in DC-MAP
Catch & Release	No **	Partially	Partially	No - but data requested by NASCO	Include collection in DC-MAP
cpue dataseries	Partially **	Partially	Partially	Partially	Improve sampling intensity in DC-MAP

Type of data	Collected under DCF	Available to WG	Reviewed and evaluated by WG	Used in current assessment models	Future plans
Age composition	Partially ** Some ageing based on fish lengths or weights	Yes	Yes	Yes	Improve coverage and sampling intensity in DC-MAP
Wild/reared origin (scale reading)	No **	Partially from other sources	Partially	Partially used - information on farmed fish is requested by NASCO	Improve sampling intensity in DC-MAP
Length and weight-at-age	Partially **	Partially	Yes	Yes - but some ageing based on fish lengths or weights	Improve sampling coverage in DC-MAP
Sex ratio	No **	Yes- from other sources	Partially	Yes	Modify sampling intensity in DC-MAP

Type of data	Collected under DCF	Available to WG	Reviewed and evaluated by WG	Used in current assessment models	Future plans
Maturity	Not known **	No *	No	No	No need to be collected – all returning adults are mature
Fecundity	No **	Yes	Partially	Yes	Include collection in DC-MAP
Data processing industry	No **	No **	No	No	No need to be collected
Juvenile surveys (Electrofishing)	Partially ** but not requested for Atlantic salmon in DCF	Yes	Partially	Partially	Include collection in DC-MAP
Adult census data (Counters, fish ladders, etc.)	Partially ** but not requested for Atlantic salmon in DCF	Yes	Partially	Yes	Include collection in DC-MAP



Add other data type to the cells with a light blue shading, if needed.

* Not asked for by the ICES WGNAS.

**) Not mandatory for some or all areas/stocks/fisheries under the current DCF.

Appendix 3: Input data for NEAC Pre Fishery Abundance analysis using Monte Carlo simulation

Finland

Annual input data for NEAC PFA run-reconstruction & NCL models for FINLAND (Uncertainty values define uniform distribution around estimates used in Monte
Carlo simulation]

Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	• •	rate (%) - MSW	. ,
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	8,422	8,538	35	5	35	5	50	10	55	15
1972	32,789	8,950	35	5	35	5	50	10	55	15
1973	15,261	14,402	35	5	35	5	50	10	55	15
1974	21,057	24,508	35	5	35	5	50	10	55	15
1975	25,242	31,347	35	5	35	5	50	10	55	15
1976	23,000	24,561	35	5	35	5	50	10	55	15
1977	12,958	17,035	35	5	35	5	50	10	55	15
1978	12,338	8,670	35	5	35	5	50	10	55	15
1979	11,071	7,078	35	5	35	5	50	10	45	15
1980	10,097	7,994	25	5	25	5	50	10	45	15
1981	9,049	9,476	25	5	25	5	50	10	45	15
1982	5,379	12,628	25	5	25	5	50	10	45	15
1983	13,156	14,013	25	5	25	5	50	10	45	15
1984	14,371	11,718	25	5	25	5	50	10	45	15
1985	19,058	11,299	25	5	25	5	50	10	45	15
1986	15,005	9,320	25	5	25	5	50	10	45	15
1987	18,151	12,208	25	5	25		50	10	45	15
1988	10,676	8,631	25	5	25	5	50	10	45	15
1989	27,956	10,337	25	5	25	5	60	10	55	15
1990	27,955	11,423	25	5	25	5	60	10	55	15
1991	27,513	15,287	25	5	25	51	60	10	55	15
1992	38,843	14,826	25	5	25	5	60	10	55	15
1993	26,195	15,517	25 25	5	25	5	60	10	55	15 15
1994	14,555	14,621	25	5			60	10	55 55	15
1995 1996	14,525	9,625	25	5	25	5	60 50	10 10	45	15
1996 1997	20,466	8,079	25	5	25	5	50	10		
1997	18,621 23,336	9,764 9,307	25	s م	25	5	50	10	45 45	15 15
1998	37,495	9,307 11,071	25	5	25	5	60	10	43 50	10
2000	40,730	21,088	25	5	25	5	60	10	50	10
2000	29,501	21,000	25		25	5	60	10	50	10
2001	16,721	26,112	25		25	5	50	10	50	10
2002	16,721	17,751	25	5	25	5	50	10	50	10
2003	7,002	8,062	25	5	25	5	50	10	50	10
2004	15,366	6,685	25	5	25	5	50	10	50	10
2005	26,916	10,533	25	5	25	5	50	10	50	10
2000	7,862	15,269	25	5	25	5	50	10	50	10
2007	8,481	15,355	25	5	25	5	50	10	50	10
2008	15,042	6,587	25	5	25	5	50	10	50	10
2009	12,085	10,590	25	5	25	5	50	10	50	10
2010	12,085	8,152	25	5	25	5	50	10	50	10
2011	23,764	9,851	25	5	25	5	50	10	50	10
2012	13,724	9,494	25	5	25	5	50	10	50	10
2013	19,495	10,302	25	5	25	5	50	10	50	10
2014	12,127	9,905	25	5	25	5	50	10	50	10
2015	9,470	10,584	25	5	25	5	50	10	50	10

France

Annual input data for NEAC PFA run-reconstruction & NCL models for FRANCE. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

Uncertainty exploitati	Estimated exploitation	rtainty in ploitation		Estimated exploitation	Uncertainty in % unreported	Estimated % unreported	Uncertainty in % unreported	Estimated % unreported	Declared catch MSW	Declared catch 1SW	Year
	rate (%) - MSW			rate (%) - 1SW	catch of MSW	catch of MSW	catch of 1SW	catch of 1SW	salmon	salmon	
salm	salmon	salmon		salmon	salmon	salmon	salmon	salmon	30111011	Sannon	
1	37.5	1.5		3.5	NA	NA	NA	NA	4,060	1,740	1971
1	37.5	1.5	;	3.5	NA	NA	NA	NA	8,120	3,480	1972
1	37.5	1.5	;	3.5	NA	NA	NA	NA	4,970	2,130	973
1	37.5	1.5	;	3.5	NA	NA	NA	NA	2,310	990	1974
1	37.5	1.5	;	3.5	NA	NA	NA	NA	4,620	1,980	1975
1	37.5	1.5	;	3.5	NA	NA	NA	NA	3,380	1,820	1976
1	37.5	1.5	;	3.5	NA	NA	NA	NA	2,600	1,400	1977
1	37.5	1.5	;	3.5	NA	NA	NA	NA	2,665	1,435	1978
1	37.5	1.5	5	3.5	NA	NA	NA	NA	3,055	1,645	1979
1	37.5	1.5	;	3.5	NA	NA	NA	NA	6,370	3,430	1980
	35	1.5		3.5	NA	NA	NA	NA	4,080	2,720	1981
	35	1.5		3.5	NA	NA	NA	NA	2,520	1,680	1982
	35	1.5		3.5	NA	NA	NA	NA	2,700	1,800	1983
	35	1.5		3.5	NA	NA	NA	NA	4,440	2,960	1984
	35	1.5		3.5	NA	NA	NA	NA	3,330	1,100	1985
	35	5		7	NA	NA	NA	NA	3,400	3,400	1986
	35	5		7	NA	NA	NA	NA	1,806	6,013	1987
	35	5		7	NA	NA	NA	NA	4,964	2,063	1988
	35	5		7	NA	NA	NA	NA	2,282	1,124	1989
	35	5		7	NA	NA	NA	NA	2,332	1,886	1990
	35	5			NA	NA	NA	NA	2,125	1,362	1991
	35	5			NA	NA	NA	NA	2,671	2,490	1992
	35	5		7	NA	NA	NA	NA	1,254	3,581	1993
	30	5		7 12.5	NA	NA	NA	NA	2,290	2,810	1994
	30	7.5 7.5		12.5	NA	NA	NA	NA	1,095 1,943	1,669 2,063	1995
	30	7.5		12.5	NA	NA NA	NA	NA	1,943	1,060	1996 1997
	30 30	7.5		12.5	NA NA	NA	NA NA	NA NA	846	2,065	1998
	30	7.5		12.5	NA	NA	NA	NA	1,831	690	1998
	30	7.5		12.5	NA NA	NA	NA	NA	1,031	1,792	2000
	30	7.5		12.5	NA	NA NA	NA	NA	1,277	1,544	2000
	30	7.5		12.5	7.5	22.5	10	30	1,065	2,423	2002
	30	7.5		12.5	7.5	22.5	10	30	1,540	1,598	2002
	30	7.5		12.5	7.5	22.5	10	30	2,880	1,927	2003
	30	7.5		12.5	7.5	22.5	10	30	1,771	1,256	2005
	30	7.5		12.5	7.5	22.5	10	30	1,785	1,763	2006
	30	7.5		12.5	7.5	22.5	10	30	1,685	1,378	2007
	30	7.5		12.5	7.5	22.5	10	30	1,865	1,365	2008
	30	7.5		12.5	7.5	22.5	10	30	863	389	2009
	30	7.5		12.5	7.5	22.5	10	30	711	1,313	2010
	30	7.5		12.5	7.5	22.5	10	30	1,998	899	2011
	30	7.5		12.5	7.5	22.5	10	30	1,585	974	2012
	30	7.5		12.5	7.5	22.5	10	30	1,632	1,371	2013
	30	7.5		12.5	7.5	22.5	10	30	2,027	1,217	2014
	30	7.5	;	12.5	7.5	22.5	10	30	2,286	1,124	2015
	30	7.5	;	12.5	7.5	22.5	10	30	972	1,017	2016

Iceland (South and West)

Annual input data for NEAC PFA run-reconstruction & NCL models for ICELAND (SOUTH-WEST). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitatio
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW		rate (%) - MSW	
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmo
1971	30,618	16,749	2	1	2	1	50	10	70	
1972	24,832	25,733	2	1	2	1	50	10	70	
1973	26,624	23,183	2	1	2	1	50	10	70	
1974	18,975	20,017	2	1	2	1	50	10	70	
1975	29,428	21,266	2	1	2	1	50	10	70	
1976	23,233	18,379	2	1	2	1	50	10	70	
1977	23,802	17,919	2	1	2	1	50	10	70	
1978	31,199	23,182	2	1	2	1	50	10	70	
1979	28,790	14,840	2		2		50	10	70	
1980	13,073	20,855		1		1	50	10	70	
1981 1982	16,890	13,919	2	1	2	1	50 50	10 10	70 70	
1982	17,331 21,923	9,826 16,423	2	1	2	1	50	10	70	
1983 1984	13,476	16,423	2	1	2	1	50	10	70	
1984 1985		13,923	2	1	2	1	50	10	70	
	21,822		2							
1986 1987	35,891 22,302	8,423 7,480	2	1	2	1	50 50	10	70 70	
1987	40,028	7,480 8,523	2	1	2	1	50	10	70	
1989	22,377	8,525 7,607	2	1	2	1	50	10	70	
1989	22,577	7,548	2	1	2	1	50	10	70	
1990	20,584	7,548	2	1	2	1	50	10	70	
1991	26,006	8,479	2	1	2	1	50	10	70	
1993	25,479	4,155	2	1	2	-	50	10	70	
1994	20,985	6,736	2	1	2	1	50	10	70	
1995	25,371	6,777	4	1	4	1	50	10	70	
1996	21,913	4,364	4	1	4	1	50	10	70	
1997	16,007	4,910	4	1	4	1	50	10	70	
1998	21,900	3,037	4	1	4	1	50	10	70	
1999	17,448	5,757	4	1	4	1	49	10	68	
2000	15,502	1,519	4	1	4	1	49	10	66	
2000	13,586	2,707	4	1	4		45	10	67	
2002	16,952	2,845	4	1	4		48	10	65	
2003	20,271	4,751	4	1	4	1	48	10	68	
2004	20,319	3,784	4	1	4		48	10	67	
2005	29,969	3,241	4	1	4	1	48	10	65	
2006	21,153	2,689	4	1	4	1	48	10	65	
2007	23,728	1,679	4	1	4	1	47	9	66	
2008	28,774	1,659	4		4	1	47	10	57	
2009	33,190	2,838	4	1	4	1	48	10	63	
2010	33,318	6,061	4	1	4	1	47	10	65	
2011	23,436	2,934	4	1	4	1	47	10	62	
2012	13,312	1,429	4	1	4	1	47	10	53	
2013	39,637	4,105	4	1	4	1	47	10	55	
2014	9,551	2,281	4	1	4	1	46	10	50	
2015	26,082	2,197	4	1	4	1	45	10	53	
	16,473	2,976	4	1	4	1	45	10	47	

Iceland (North and East)

Annual input data for NEAC PFA run-reconstruction & NCL models for ICELAND (NORTH-EAST). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Uncertainty exploitation	Estimated exploitation	Uncertainty in exploitation	Estimated exploitation	Uncertainty in % unreported	Estimated % unreported	Uncertainty in % unreported	Estimated % unreported	Declared catch MSW	Declared catch 1SW	Year
	rate (%) - MSW		rate (%) - 1SW	catch of MSW	catch of MSW	catch of 1SW	catch of 1SW	salmon	salmon	
salm	salmon	salmon	salmon	salmon	salmon	salmon	salmon	Sumon	Sannon	
54111	70	10	50	1	2	1	2	6,625	4,610	1971
	70	10	50	1	2	1	2	10,337	4,223	972
	70	10	50	1	2	1	2	9,672	5,060	973
	70	10	50	1	2	1	2	9,176	5,047	1974
	70	10	50	1	2	1	2	10,136	6,152	1975
	70	10	50	1	2	1	2	8,350	6,184	1976
	70	10	50	1	2	1	2	11,631	8,597	1977
	70	10	50	1	2	1	2	14,998	8,739	1978
	70	10	50	1	2	1	2	9,897	8,363	1979
	70	10	50	1	2	1	2	13,784	1,268	1980
	70	10	50	1	2	1	2	4,827	6,528	1981
	70	10	50	1	2	1	2	5,539	3,007	1982
	70	10	50	1	2	1	2	4,224	4,437	1983
	70	10	50	1	2	1	2	5,447	1,611	L984
	70	10	50	1	2	1	2	3,511	11,116	1985
	70	10	50	1	2	1	2	9,569	13,827	1986
	70	10	50	1	2	1	2	9,908	8,145	1987
	70	10	50	1	2	1	2	6,381	11,775	1988
	70	10	50	1	2	1	2	5,414	6,342	1989
	70	10	50	1	2	1	2	5,709	4,752	1990
	70	10	50	1	2	1	2	3,965	6,900	1991
	70	10	50	1	2	1	2	5,903	12,996	992
	70	10	50	1	2	1	2	6,672	10,689	1993
	70	10	50	1	2	1	2	5,656	3,414	1994
	70	10	50	1	4	1	4	3,511	8,776	1995
	70	10	50	1	4	1	4	4,605	4,681	1996
	70	10	50	1	4	1	4	2,594	6,406	1997
	70	10	50	1	4	1	4	3,780	10,905	1998
	65	10	48	1	4	1	4	4,030	5,326	1999
	64	10	48	1	4	1	4	2,324	5,595	2000
	62 60	10 10	47 46		4	1	4	2,587	4,976 8,437	2001 2002
	53	10	46	1	4	1	4	2,366	8,437 4,478	2002
	55	10	46	1	4	1	4	2,194 2,239	4,478	2003
	54	10	43	1	4	1	4	2,239	10,297	2004
	45	10	44	1	4	1	4	2,720	10,297	2005
	36	10	45	1	4	1	4	1,672	8,046	2007
	45	10	44	1	4		4	2,693	7,021	2008
	36	10	42	1	4		4	1,735	10,779	2009
	38	10	40	1	4	1	Å	2,602	8,621	2010
	34	10	38	1	4	1	4	2,596	6,759	2011
	33	10	40	1	4	1		1,419	3,699	2012
	31	10	38	1	4	1	. 4	1,528	8,375	2013
	30	10	38	1	4	1	4	1,778	3,953	2014
	32	10	35	1	4	1	4	1,803	10,209	2015
	44	10	44	1	4	1		3,361	7,870	2015

Annual input data for NEAC PFA run-reconstruction & NCL models for IRELAND. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Declared		,	Estimated % Un			Uncertainty in		,	Declared net		Catch and		1SW salmon MS			MSW salmon
	catch 1SW	catch MSW		% unreported						exploitation	catch 1SW	catch MSW	release 1SW	release MSW	in Small rivers in Sr	mall rivers	in closed	in closed
	salmon	salmon	catch of 1SW salmon	catch of 1SW salmon	catch of MSW cat salmon	tch of MSW salmon	rate (%) - 1SW salmon	. ,	rate (%) - MSW salmon	rate (%) - MSW salmon	salmon	salmon	salmon	salmon			rivers	river
1971	409,965	46,594	37.5	7.5		7.5	62.5	12.5	47.5	12.5	NA	NA	NA	NA	NA	NA	NA	NA
1972	437,089	49,863	37.5			7.5	62.5	12.5	47.5	12.5		NA	NA	NA	NA	NA	NA	NA
1973	476,131	54,008	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5	NA_	NA	NA	NA	NA	NA	NA	N
1974	542,124	60,976	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5	NA	NA	NA	NA	NA	NA	NA	N
1975	598,524	68,260	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5	NA	NA	NA	NA	NA	NA	NA	N
1976	407,018	47,358	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5	NA	NA	NA	NA	NA	NA	NA	NA
1977	351,745	41,256	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5	NA	NA	NA	NA	NA	NA	NA	NA
1978	307,569	35,708	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5	NA	NA	NA	NA	NA	NA	NA	NA
1979	282,700	32,144	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5		NA	NA	NA	NA	NA	NA	NA
1980	215,116	35,447	37.5		37.5	7.5	62.5	12.5	47.5	12.5	NA	NA	NA	NA	NA	NA	NA	NA
1981	137,366	26,101	37.5	7.5	37.5	7.5	75.7	11.4	47.5	12.5	NA	NA	NA	NA	NA	NA	NA	NA
1982	269,847	11,754	37.5	7.5	37.5	7.5	71.9	10.8	36.7	8.3	NA	NA	NA	NA	NA	NA	NA	NA
1983	437,751	26,479	37.5		37.5	7.5	66.1	9.9	40.1	7.5	NA	NA	NA	NA	NA	NA	NA	NA
1984	224,872	20,685	37.5		37.5	7.5	64.6	9.7	43.5	6.5		NA	NA	NA	NA	NA	NA	NA
1985	430,315	18,830	37.5			7.5	74.6	11.2	36.1	3.4	NA	NA	NA	NA	NA	NA	NA	NA
1986	443,701	27,111	37.5	7.5	37.5	7.5	68.7	10.3	46.0	9.0	NA	NA	NA	NA	NA	NA	NA	NA
1987	324,709	26,301	30			10	69.8	10.5	32.2	4.7	NA	NA	NA	NA	NA	NA	NA	NA
1988	391,475	22,067	30	10	30	10	62.0	9.3	37.4			NA	NA	NA	NA	NA	NA	NA
1989	297,797	25,447	30	10	30	10	65.7	9.9	47.2	8.8		NA	NA	NA	NA	NA	NA	NA
1990	172,098	15,549	30	10	30	10	60.7	9.1	59.9	6.1	NA	NA	NA	NA	NA	NA	NA	NA
1991 1992	120,408 182,255	10,334	30 30	10 10	30 30	10 10	59.5 62.1	8.9 9.3	26.5 51.5	3.5 3.8		NA	NA	NA	NA NA	NA NA	NA	NA
1992	182,255	15,456 13,156	25	10	25	10	58.6	9.3		3.8 18.0		NA NA	NA NA	NA NA	NA	NA	NA NA	NA NA
1993	234,126	20,506	25	10		10	71.4	8.8 10.7	42.0	2.5		NA	NA	NA	NA	NA	NA	N/
1995	234,120	20,300	25	10		10	63.5	10.7	40.5	1.2		NA	NA	NA	NA	NA	NA	N/
1995	203,920	18,021	25			10	59.9	9.0		3.2		NA	NA	NA	NA	NA	NA	NA
1997	170,774	14,724	25	10		5	50.1	7.5		12.2		NA	NA	NA	NA	NA	NA	NA
1998	191,868	17,269	25	10		5	53.7	8.1	61.9	1.4		NA	NA	NA	NA	NA	NA	NA
1999	158,818	14,801	25	10		5	47.8			18.1		NA	NA	NA	NA	NA	NA	NA
2000	199,827	16,848	25			5	43.2	6.5		4.5		NA	NA	NA	NA	NA	NA	NA
2001	218,715	18,436	7.5		7.5	2.5	48.0	7.2		8.0		NA	NA	NA	NA	NA	NA	NA
2002	198,719	16,702	7.5		7.5	2.5	49.9	7.5		7.5	NA	NA	NA	NA	NA	NA	NA	NA
2003	161,270	13,745	7.5	2.5	7.5	2.5	41.3	6.2	21.5	5.5	NA	NA	NA	NA	NA	NA	NA	NA
2004	142,251	12,299	7.5	2.5	7.5	2.5	49.5	7.5	35.0	8.0	NA	NA	NA	NA	NA	NA	NA	NA
2005	127,371	10,716	7.5	2.5	7.5	2.5	44.5	6.5	23.5	3.5	NA	NA	NA	NA	NA	NA	NA	NA
2006	101,938	9,740	7.5	2.5	7.5	2.5	46.5	6.5	29.5	13.5	NA	NA	NA	NA	NA	NA	NA	NA
2007	17,863	2,867	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	8,177	666	12,137	988	0	0	24,433	158
2008	31,843	3,935	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	8,233	670	12,071	1,492	0	0	23,259	213
2009	24,268	4,675	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	6,248	509	9,812	1,610	0	0	30,008	1,873
2010	32,981	4,497	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	-,	1,066	13,325	1,817	0	0	30,605	616
2011	28,105	4,889	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	,	902	11,031	1,657	0	0	28,504	765
2012	29,979	4,197	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	- / -	777	10,429	1,463	0	0	24,517	1,213
2013	24,029	4,831	7.5		7.5	2.5	15.5	8.4		9.1		747	8,821	1,861	0	0	23,836	1,250
2014	13,787	4,063	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1		397	5,107	1,430	0	0	20,110	1,210
2015	20,835	4,272	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1		295	7,810	1,573	0	0	25,834	1,147
2016	21,084	3,783	7.5	2.5	7.5	2.5	15.5	8.4	23.9	9.1	5,755	1,162	8,890	1,391	0	0	23,952	1,657

Norway (Southeast)

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (SOUTH-EAST). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Uncertaint [®] exploitat	Estimated exploitation	Uncertainty in exploitation	Estimated exploitation	Uncertainty in % unreported	Estimated % unreported	Uncertainty in % unreported	Estimated % unreported	Declared catch MSW	Declared catch 1SW	Year
	rate (%) - MSW		rate (%) - 1SW	catch of MSW	catch of MSW	catch of 1SW	catch of 1SW	salmon	salmon	
salm	salmon	salmon	salmon	salmon	salmon	salmon	salmon	Sannon	samon	
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1971
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1972
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1973
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1974
	NA	NA	NA	NA	NA	NA	NA	NA	NA	1975
	NA	NA	NA	NA	NA	NA	NA	NA	NA	1976
	NA	NA	NA	NA	NA	NA	NA	NA	NA	1977
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1978
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1979
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1980
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1981
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1982
	65	10	70	10	50	10	50	9,004	9,039	1983
	65	10	70	10	50	10	50	11,527	11,402	1984
	65	10	70	10	50	10	50	11,883	18,699	1985
	65	10	70	10	50	10	50	12,077	23,089	1986
	65	10	70	10	50	10	50	14,179	19,601	1987
	65	10	70	10	50	10	50	9,443	17,520	1988
	60	10	65	10	50	10	50	12,254	23,965	1989
	60	10	65	10	50	10	50	11,502	25,792	1990
	60	10	65	10	50	10	50	10,753	21,064	1991
	60	10	65	10	50	10	50	15,332	26,044	1992
	60	10	65	10	40	10	40	12,596	23,070	1993
	60	10	65	10	40	10	40	9,988	23,987	1994
	60	10	65	10	40	10	40	11,630	21,847	1995
	60	10	65	10	40	10	40	13,538	20,738	1996
	60	10	60	10	35	10	35	7,756	21,121	1997
	60	10	60	10	35	10	35	10,396	32,586	1998
	60	10	60	10	35	10	35	6,664	23,904	1999
	60	10	60	10	35	10	35	14,261	43,151	2000
	60	10	60	10	35	10	35	19,210	47,339	2001
	60	10	60	10	35	10	35	14,400	33,087	2002
	60	10	60	10	30	10	30	20,648	33,371	2003
	60	10	60	10	30	10	30	15,948	28,506	2004
	60	10	60	10	30	10	30	14,628	40,628	2005
	60	10	60	10	30	10	30	21,192	30,979	2006
	60	10	60	10	30	10	30	18,130	15,735	2007
	50	10	55	10	30	10	30	16,678	15,696	2008
	50	10	55	10	30	10	30	11,995	15,584	2009
	40	10	50	10	30	10	30	12,175	22,139	2010
	40	10	50	10	30	10	30	28,589	15,773	2011
	40	10	50	10	30	10	30	23,389	18,582	2012
	40	10	50	10	30	10	30	13,564	16,702	2013
	35	10	40	10	30	10	30	13,699	15,389	2014
	40	10	50	10	30	10	30	17,079	17,188	2015
	40	10	50	10	30	10	30	18,579	14,670	2016

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (SOUTH-WEST) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

	Uncertainty in % unreported	Estimated % unreported	Uncertainty in % unreported	Estimated % unreported	Declared catch MSW	Declared catch 1SW	Year
	catch of MSW	catch of MSW	catch of 1SW	catch of 1SW	salmon	salmon	
salmon salmon salmon salmon salmon salmon		salmon	salmon	salmon	samon	Sannon	
NA NA NA NA		NA	NA	NA	NA	NA	1971
NA NA NA NA		NA	NA	NA	NA	NA	1972
NA NA NA NA		NA	NA	NA	NA	NA	1973
NA NA NA NA		NA	NA	NA	NA	NA	1974
NA NA NA NA		NA	NA	NA	NA	NA	1975
NA NA NA NA	NA	NA	NA	NA	NA	NA	1976
NA NA NA NA		NA	NA	NA	NA	NA	1977
NA NA NA NA	NA	NA	NA	NA	NA	NA	1978
NA NA NA NA	NA	NA	NA	NA	NA	NA	1979
NA NA NA NA	NA	NA	NA	NA	NA	NA	1980
NA NA NA NA	NA	NA	NA	NA	NA	NA	1981
NA NA NA NA	NA	NA	NA	NA	NA	NA	1982
10 80 10 80	10	50	10	50	28,601	31,845	1983
10 80 10 80	10	50	10	50	27,641	23,428	1984
10 80 10 80	10	50	10	50	25,515	29,857	1985
10 80 10 80	10	50	10	50	30,769	29,894	1986
10 80 10 80	10	50	10	50	26,623	30,005	1987
10 80 10 80	10	50	10	50	28,255	36,976	1988
10 70 10 65	10	50	10	50	13,041	19,183	1989
10 70 10 65	10	50	10	50	14,423	18,490	1990
10 70 10 65	10	50	10	50	8,323	9,759	1991
10 70 10 65		50	10	50	8,832	6,448	1992
10 70 10 65	10	40	10	40	10,239	11,433	1993
10 70 10 65		40	10	40	10,961	18,597	1994
10 70 10 65	10	40	10	40	13,122	10,863	1995
10 70 10 65		40	10	40	12,546	7,048	1996
10 60 10 60		35	10	35	7,194	10,279	1997
10 60 10 60		35	10	35	6,583	5,726	1998
10 60 10 60		35	10	35	3,219	7,357	1999
10 60 10 60		35	10	35	7,961	11,538	2000
10 60 10 60		35	10	35	10,716	12,109	2001
10 60 10 60		35	10	35	7,145	6,000	2002
10 60 10 60		30	10	30	7,602	8,269	2003
10 60 10 60		30	10	30	6,420	7,180	2004
10 60 10 60		30	10	30	7,334	10,370	2005
10 60 10 60		30 30	10	30	9,381	5,173	2006
10 60 10 60		30	10	30 30	6,011	2,630	2007 2008
10 55 10 50 10 55 10 50		30	10	30 30	4,807 3,792	3,143 3,069	2008
10 50 10 35		30	10	30	2,447	3,450	2009
10 45 10 30		30	10	30	4,409	2,888	2010
10 45 10 30 10 45 10 30		× 30 30	10	30	4,409 5,733	2,888 4,171	2011
10 45 10 50 10 45 10 30		30	10	30	3,581	4,171 3,111	2012
10 40 10 25		30	10	30	2,717	3,029	2013
	10	30	10	30	3,953	4,721	2014
10 45 10 30	10	30	10	30	5,671	3,262	2016

Mid-Norway

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (MID AREA) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

exploitatio	Estimated exploitation	Uncertainty in exploitation	Estimated exploitation	Uncertainty in % unreported	Estimated % unreported	Uncertainty in % unreported	Estimated % unreported	Declared catch MSW	Declared catch 1SW	Year
	rate (%) - MSW		rate (%) - 1SW	catch of MSW	catch of MSW	catch of 1SW	catch of 1SW	salmon	salmon	
salmo	salmon	salmon	salmon	salmon	salmon	salmon	salmon	Sannon	Sannon	
N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1971
N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1972
N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1973
N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1974
N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1975
N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1976
N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1977
N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1978
Ν	NA	NA	NA	NA	NA	NA	NA	NA	NA	1979
Ν	NA	NA	NA	NA	NA	NA	NA	NA	NA	1980
N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1981
N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1982
	75	10	75	10	50	10	50	74,648	121,221	1983
	75	10	75	10	50	10	50	67,639	94,373	1984
	75	10	75	10	50	10	50	56,641	114,613	1985
	75	10	75	10	50	10	50	77,225	106,921	1986
	75	10	75	10	50	10	50	62,216	83,669	1987
	75	10	75	10	50	10	50	45,609	80,111	1988
:	65	10	65	10	50	10	50	30,862	94,897	1989
:	65	10	65	10	50	10	50	40,174	78,888	1990
	65	10	65	10	50	10	50	30,087	67,370	1991
	65	10	65	10	50	10	50	33,092	51,463	1992
	65	10	65	10	40	10	40	28,184	58,326	1993
	65	10	65	10	40	10	40	33,520	113,427	1994
	65	10	65	10	40	10	40	42,696	57,813	1995
	65	10	65	10	40	10	40	31,613	28,925	1996
	60	10	60	10	35	10	35	20,565	43,127	1997
	60	10	60	10	35	10	35	26,817	63,497	1998
	60	10	60	10	35	10	35	28,792	60,689	1999
	60	10	60	10	35	10	35	42,452	109,278	2000
	60	10	60	10	35	10	35	52,031	88,096	2001
	60	10	60	10	35	10	35	52,774	42,669	2002
	60	10	60	10	30	10	30	46,963	91,118	2003
	60	10	60	10	30	10	30	49,760	38,286	2004
	60	10	60	10	30	10	30	37,941	63,749	2005
	60	10	60	10	30 30	10	30	47,691	46,495	2006
	60	10	60	10	'	10	30	33,106	26,608	2007
	45	10	55	10	30	10 10	30	34,869	31,936	2008
	45	10	55	10	30		30	30,715	26,267	2009
	45	10	50	10	30	10	30	30,524	37,557	2010
	45	10	50 50	10	30	10	30 30	37,272	20,932	2011
	45 40	10 10	45	10 10	30 30	10 10	30	28,265 17,727	22,368 25,121	2012 2013
	40 32	10	45	10	30	10	30	17,727		2013
	32 40	10	45	10	30	*	30 30	30,457	25,349 30,932	2014
		10	40	10	30	10	30	26,325	20,498	2015

Norway North

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (NORTH) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

exploitat	Estimated exploitation	Uncertainty in exploitation	Estimated exploitation	Uncertainty in % unreported	Estimated % unreported	Uncertainty in % unreported	Estimated % unreported	Declared catch MSW	Declared catch 1SW	Year
	rate (%) - MSW	•	rate (%) - 1SW	catch of MSW	catch of MSW	catch of 1SW	catch of 1SW	salmon	salmon	
salm	salmon	salmon	salmon	salmon	salmon	salmon	salmon	Sdillioli	SdiiiiOli	
Sain N	NA	NA	NA	NA	NA	NA	NA	NA	NA	1971
ר	NA	NA	NA	NA	NA	NA	NA	NA	NA	1972
r	NA	NA	NA	NA	NA	NA	NA	NA	NA	.973
r	NA	NA	NA	NA	NA	NA	NA	NA	NA	1974
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1975
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1976
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1977
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1978
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1979
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1980
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1981
1	NA	NA	NA	NA	NA	NA	NA	NA	NA	1982
	80	10	80	10	50	10	50	49,413	104,040	1983
	80	10	80	10	50	10	50	58,858	150,372	1984
	80	10	80	10	50	10	50	58,956	118,841	1985
	80	10	80	10	50	10	50	63,418	84,150	1986
	80	10	80	10	50	10	50	34,232	72,370	1987
	80	10	80	10	50	10	50	32,140	53,880	1988
	70	10	70	10	50	10	50	13,934	42,010	1989
	70	10	70	10	50	10	50	17,321	38,216	1990
	70	10	70	10	50	10	50	21,789	42,888	1991
	70	10	70	10	50	10	50	19,265	34,593	1992
	70	10	70	10	40	10	40	39,014	51,440	1993
	70	10	70	10	40	10	40	33,411	37,489	1994
	70	10	70	10	40	10	40	26,037	36,283	1995
	70	10	70	10	40	10	40	36,636	40,792	1996
	70	10	70	10	35	10	35	30,115	39,930	1997
	70	10	70	10	35	10	35	34,806	46,645	1998
	70	10	70	10	35	10	35	46,744	46,394	1999
	70	10	70	10	35	10	35	51,569	61,854	2000
	70	10	70	10	35	10	35	54,023	46,331	2001
	70	10	70	10	35	10	35	43,100	38,101	2002
	70	10	70	10	30	10	30	35,972	44,947	2003
	70	10	70	10	30	10	30	28,077	34,640	2004
	70	10	70	10	30	10	30	33,334	45,530	2005
	70 70	10 10	70 70	10 10	30 30	10 10	30 30	39,508	48,688 28,748	2006 2007
		10	65		30		30	44,550		2007
	65 65	10	65	10 10	30	10 10	30	40,553	34,338	2008
	55	10	65	10	30	10	30	28,241 28,611	22,511 29,836	2009
		10		10	*		30			
	55 55	10	65 65	10	30 30	10 10	30	27,233 28,000	26,813 28,289	2011 2012
	55	10	65	10	30	10	30 30	28,000	28,289 20,021	2012
	55	10	65	10	30	10	30	24,089	35,171	2015
	55	10	65	10	30	10	30	23,810	25,426	2014
	55	10	65	10	30	10	30	33,607	23,420	2015

Russia (Archangelsk and Karelia)

Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (ARCHANGLE/KORELIA). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

Uncertainty exploitati	Estimated exploitation	Uncertainty in exploitation	Estimated exploitation	Uncertainty in % unreported	Estimated % unreported	Uncertainty in % unreported	Estimated % unreported	Declared catch MSW	Declared catch 1SW	Year
	rate (%) - MSW		rate (%) - 1SW	catch of MSW	catch of MSW	catch of 1SW	catch of 1SW	salmon	salmon	
salm	salmon	salmon	salmon	salmon	salmon	salmon	salmon	Sumon	Sannon	
	60	20	60	5	10	5	10	16,592	134	1971
	60	20	60	5	10	5	10	14,434	116	1972
	60	20	60	5	10	5	10	20,924	169	1973
	60	20	60	5	10	5	10	21,137	170	1974
	60	20	60	5	10	5	10	17,398	140	1975
	60	20	60	5	10	5	10	13,781	111	1976
	60	20	60	5	10	5	10	9,722	78	1977
	60	20	60	5	10	5	10	10,134	82	1978
	60	20	60	5	10	5	10	13,903	112	1979
	60	20	60	5	10	5	10	19,397	156	1980
	60	20	60	5	10	5	10	8,394	68	1981
	60 60	20	60 60	5	10	5	10	8,797	71	1982
	60	20 20	60	5	10	5	10	11,938	48 21	1983 1984
	60	20	60	5	10 10	5	10 10	10,680 11,183	454	1984 1985
	60	20	60	5	10	5	10	12,291	434	1985
	60	20	60	5	10	5	10	8,734	647	1987
	60	20	60	5	10	5	10	9,978	224	1988
	60	20	60	5	10	5	10	10,245	989	1989
	60	20	60	5	15	5	15	8,429	1,418	1990
	60	20	60	5	20	5	20	8,725	421	1991
	60	20	60	5	25	5	25	3,949	1,031	1992
	60	20	60	5	30	5	30	4,251	196	1993
	60	20	60	5	35	5	35	5,631	334	1994
	60	20	60	5	45	5	45	5,214	386	1995
	60	20	60	5	55	5	55	3,753	231	1996
	60	20	60	5	55	5	55	3,351	721	1997
	60	20	60	5	55	5	55	4,208	585	1998
	60	20	60	5	55	5	55	3,101	299	1999
	60	20	60	5	55	5	55	3,382	514	2000
	60	20	60	5	55	5	55	2,348	363	2001
	60	20	60	5	55	5	55	2,439	1,676	2002
	60	20	60	5	55	5	55	2,041	893	2003
	60	20	60	5	55	5	55	3,761	990	2004
	60	20	60	5	55	5	55	4,915	1,349	2005
	60	20	60	5	55 55	5	55	2,841	2,183	2006
	60 60	20 20	60 60	5	55	5	55 55	2,621	1,618 332	2007 2008
	60	20	60	5	55	5	55	2,496 2,214	252	2008
	60	20	60	5	55	5	55	3,823	397	2009
	60	20	60	5	55	5	55	2,585	313	2010
	60	20	60	5	55	5	55	2,565	1,332	2011
	60	20	60	5	55	5	55	3,480	2,296	2012
	60	20	60	5	55	5	55	3,463	2,084	2014
	60	20	60	5	55	5	55	3,542	2,071	2015
	60	20	60	5	55	5	.55	2,221	3,042	2016

Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (KOLA-BARENT SEA). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Uncertainty exploitation	Estimated exploitation	Uncertainty in exploitation	Estimated exploitation	Uncertainty in % unreported	Estimated % unreported	Uncertainty in % unreported	Estimated % unreported	Declared catch MSW	Declared catch 1SW	Year
•	rate (%) - MSW		rate (%) - 1SW	catch of MSW	catch of MSW		catch of 1SW	salmon	salmon	
salmo	salmon	salmon	salmon	salmon	salmon	salmon	salmon	Sannon	Samon	
Surrie	45	5	45	5	15	5	15	5,979	4,892	1971
	45	5	45	5	15	5	15	9,750	7,978	972
	40	5	40	5	15	5	15	11,460	9,376	1973
	40	5	40	5	15	5	15	15,638	12,794	1974
	45	5	45	5	15	5	15	13,872	13,872	1975
	55	5	55	5	15	5	15	14,048	11,493	1976
	50	5	50	5	15	5	15	8,253	7,257	1977
	55	5	55	5	15	5	15	7,113	7,106	1978
	40	5	40	5	15	5	15	3,141	6,707	1979
	40	5	40	5	15	5	15	5,216	6,621	1980
	40	5	40	5	15	5	15	5,973	4,547	1981
	35	5	35	5	15	5	15	4,798	5,159	1982
	35	5	35	5	15	5	15	9,943	8,504	1983
	35	5	35	5	15	5	15	12,601	9,453	1984
	35	5	35	5	15	5	15	7,877	6,774	1985
	40	5	40	5	15	5	15	5,352	10,147	1986
	40	5	40	5	15	5	15	5,149	8,560	L987
	35	5	35	5	15	5	15	3,655	6,644	1988
	40	5	40	5	15	5	15	6,787	13,424	1989
	40	5	40	5	15	5	15	8,234	16,038	1990
	30	5	30	5	15	5	15	7,568	4,550	1991
	30	5	30	5	15	5	15	7,109	11,394	992
	30	5	30	5	15	5	15	5,690	8,642	1993
	30	5	30	5	15	5	15	4,632	6,101	1994
	30	5	30		15	5	15	3,693	6,318	1995
	25 15	5	15	5 5	20 25	5	20 25	1,701 867	6,815 3,564	1996 1997
	15	3	13	5	25	5	35	280	3,564	1997
	13	3	8	5	40	5	40	424	1,834	1998
	6	2	6	5	40 50	5	40 50	323	805	2000
	4	2	4	3	60	5	60	241	591	2000
	20	5	10	10	50	10	50	2,478	1,436	2002
	20	5	10	10	50	10	50	1,095	1,938	2003
	20	5	10	10	50	10	50	850	1,095	2004
	20	5	10	10	60	10	60	426	859	2005
	20	5	10	10	60	10	60	844	1,372	2006
	20	5	10	10	60	10	60	707	784	2007
	20	5	15	10	60	10	60	997	1,446	2008
	20	5	15	10	60	10	60	1,080	2,882	2009
	25	5	20	10	60	10	60	1,486	3,884	2010
	25	5	20	10	60	10	60	1,407	3,861	2011
	25	5	20	10	60	10	60	1,027	2,708	2012
	25	5	20	10	60	10	60	904	939	2013
	25	5	20	10	60	10	60	789	969	2014
	25	5	20	10	60	10	60	494	727	2015
	25	5	20	10	60	10	60	625	380	2016

Russia (Kola Peninsula: White Sea Basin)

Declared	Declared	Catch 1SW	Catch MSW		Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty i
catch 1SW salmon	catch MSW salmon	following-year spawners	following-year spawners	unreported catch of 1SW	% unreported catch of 1SW	unreported catch of MSW	% unreported catch of MSW	exploitation rate (%) - 1SW	exploitation rate (%) - 1SW	exploitation rate (%) - MSW	exploitatio
			•	salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmo
67,845 45,837	29,077 19,644	0	0	3	2	3	2	50 50	10 10	60 60	1
45,837 68,684	29,436	0	0	3	2	3	2	50	10	60	1
63,892		0	0	3	2	3	2	50	10	60	1
109,038	46,730	0	0	3	2	3	2	50	10	60	1
76,281	41,075	0	0	3	2	3		50	10	60	1
											1
											1
											1
40,172	12,747	0	0	3	2	3	2	50	10	60	1
32,619	14,840	0	0	3	2	3		50	10	60	1
											1
											1
											1
		0	0	3	2	3		50	10	50	1
44,813	18,777	0	0	3	2	3	2	45	5	45	
53,293	11,448	0	0	8	3	8	3	45		45	
25,178	3,680	0	0	30	5	30	- 5	25	5	15	
19,381	2,847	0	Ō	35	5	35	4	25	5	15	
27,097	2,710	0	0	35	5	35	5	25	5	15	
							5				
		0	0	35		35	5	25		15	
20,483	3,959	0	0	35	5	35	5	15	5	15	
19,174	3,937	0	0	35	5	35	5	15	5	15	
							5				
		3.852					5				
7,807	1,404	2,264	852	35	5	35	5	15	5	15	
8,447	4,711	3,175	832	35	5	35	5	15	5	15	
					5						
8,986	1,660	2,203	734	35	5	35	5	15	5	15	
8,593	1,674	3,307	1,102	35	5	35	5	15	5	15	
					5						
	63,820 (34,25) (5,28) (47,94) (44,91)	$\begin{array}{rrrrrrrrrrrrrrrrrrrrrrrrrrrrrrrrrrrr$	$\begin{array}{c} 63,802 & 27,822 & 00 \\ 109,038 & 46,730 & 00 \\ 76,281 & 41,075 & 00 \\ 47,943 & 32,392 & 00 \\ 49,291 & 17,307 & 00 \\ 69,511 & 21,369 & 00 \\ 40,172 & 12,747 & 00 \\ 32,619 & 14,840 & 00 \\ 56,766 & 16,893 & 00 \\ 56,766 & 16,893 & 00 \\ 57,2102 & 17,681 & 00 \\ 79,639 & 12,501 & 00 \\ 79,639 & 12,501 & 00 \\ 32,327 & 3,680 & 00 \\ 32,327 & 3,680 & 00 \\ 32,327 & 3,680 & 00 \\ 32,327 & 3,680 & 00 \\ 32,327 & 3,680 & 00 \\ 32,327 & 3,680 & 00 \\ 32,327 & 3,680 & 00 \\ 32,327 & 3,680 & 00 \\ 23,327 & 3,680 & 00 \\ 23,327 & 3,680 & 00 \\ 23,327 & 3,680 & 00 \\ 22,330 & 2,841 & 00 \\ 27,695 & 2,085 & 00 \\ 22,330 & 2,841 & 00 \\ 22,633 & 1,963 & 00 \\ 22,330 & 2,841 & 00 \\ 22,637 & 4,395 & 00 \\ 33,172 & 2,388 & 1,212 \\ 15,667 & 3,734 & 00 \\ 19,174 & 3,937 & 00 \\ 19,174 & 3,937 & 00 \\ 19,174 & 3,937 & 00 \\ 19,174 & 3,937 & 00 \\ 19,174 & 3,937 & 00 \\ 19,174 & 3,937 & 00 \\ 10,947 & 1,9930 & 00 \\ 24,330 & 2,841 & 00 \\ 24,330 & 2,841 & 00 \\ 24,330 & 2,841 & 00 \\ 24,330 & 2,841 & 00 \\ 24,330 & 2,841 & 00 \\ 24,330 & 2,841 & 00 \\ 24,330 & 1,953 & 00 \\ 24,330 & 2,841 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 24,330 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 25,515 & 00 \\ 2$	$\begin{array}{c c c c c c c c c c c c c c c c c c c $	$\begin{array}{c c c c c c c c c c c c c c c c c c c $	$\begin{array}{c c c c c c c c c c c c c c c c c c c $	63,892 $27,382$ 00323 $109,038$ $46,730$ 00323 $47,943$ $32,292$ 00323 $47,943$ $32,292$ 00323 $49,291$ $17,307$ 00323 $46,037$ $22,441$ 00323 $46,037$ $12,2477$ 00323 $32,619$ $14,840$ 00323 $56,786$ $16,893$ 00323 $56,786$ $16,893$ 00323 $72,102$ $17,681$ 00323 $79,639$ $12,501$ 00323 $34,479$ 003233 $31,978$ $6,263$ 0018318 $23,827$ $3,660$ 0035535 $20,987$ $5,552$ 002555 $20,987$ $2,710$ 0035535 $27,697$ $2,710$ 0035535 $20,987$ $3,724$ 0035535 $21,939$ 0 035535 $32,633$ $1,963$ 0035535 $32,637$ $1,963$ 0035535 <t< td=""><td>63,822$27,822$$0$$0$$3$$2$$3$$2$$109,038$$46,730$$0$$0$$3$$2$$3$$2$$47,943$$32,992$$0$$0$$3$$2$$3$$2$$47,943$$32,992$$0$$0$$3$$2$$3$$2$$69,511$$21,369$$0$$0$$3$$2$$3$$2$$46,037$$23,241$$0$$0$$3$$2$$3$$2$$46,037$$23,241$$0$$0$$3$$2$$3$$2$$56,786$$16,893$$0$$0$$3$$2$$3$$2$$56,786$$16,893$$0$$0$$3$$2$$3$$2$$72,102$$17,681$$0$$0$$3$$2$$3$$2$$73,639$$12,501$$0$$0$$3$$2$$3$$2$$73,639$$12,501$$0$$0$$3$$2$$3$$2$$73,639$$12,501$$0$$0$$3$$2$$3$$2$$73,639$$12,501$$0$$0$$3$$2$$3$$3$$23,311,448$$0$$0$$3$$2$$3$$3$$3$$23,321,1,448$$0$$0$$3$$3$$3$$3$$3$$23,927$$3,660$$0$$23$$3$$23$$3$$3$$3$$3$$23,937$$0$$0$$35$</td></t<> <td>63,892 7,382 0 0 3 2 3 2 90 109,038 46,730 0 0 3 2 3 2 90 47,943 32,322 0 0 3 2 3 2 90 49,291 17,307 0 0 3 2 3 2 90 46,037 12,1369 0 0 3 2 3 2 90 46,037 12,2477 0 0 3 2 3 2 90 32,619 14,480 0 0 3 2 3 2 90 56,786 16,893 0 0 3 2 3 2 90 72,102 17,681 0 0 3 2 3 2 90 79,639 12,501 0 0 3 2 3 2 90 79,639 12,501 0 0 3 3 3 3 3 3 3<!--</td--><td>63,822 27,822 0 0 3 2 3 2 50 10 109,038 46,730 0 0 3 2 3 2 50 100 47,943 32,992 0 0 3 2 3 2 50 100 49,291 17,307 0 0 3 2 3 2 50 100 46,037 23,241 0 0 3 2 3 2 50 100 46,037 23,241 0 0 3 2 3 2 50 100 54,177 0 0 3 2 3 2 50 100 56,786 16,893 0 0 3 2 3 2 50 100 72,102 17,681 0 0 3 2 3 2 50 100 73,639 12,601 0 0 3 2 3 2 55 55 55 55 5</td><td>63,89227,382003232501060109,03346,73000323250106047,94132,38200323250106047,94332,38200323250106069,51121,36900323250106064,01722,74100323250106054,21720,84000323250106054,21720,84000323250106054,21720,84000323250106054,21720,84000323250106072,1217,68100323250106072,1217,68100323250106072,1314833133455</td></td>	63,822 $27,822$ 0 0 3 2 3 2 $109,038$ $46,730$ 0 0 3 2 3 2 $47,943$ $32,992$ 0 0 3 2 3 2 $47,943$ $32,992$ 0 0 3 2 3 2 $69,511$ $21,369$ 0 0 3 2 3 2 $46,037$ $23,241$ 0 0 3 2 3 2 $46,037$ $23,241$ 0 0 3 2 3 2 $56,786$ $16,893$ 0 0 3 2 3 2 $56,786$ $16,893$ 0 0 3 2 3 2 $72,102$ $17,681$ 0 0 3 2 3 2 $73,639$ $12,501$ 0 0 3 2 3 2 $73,639$ $12,501$ 0 0 3 2 3 2 $73,639$ $12,501$ 0 0 3 2 3 2 $73,639$ $12,501$ 0 0 3 2 3 3 $23,311,448$ 0 0 3 2 3 3 3 $23,321,1,448$ 0 0 3 3 3 3 3 $23,927$ $3,660$ 0 23 3 23 3 3 3 3 $23,937$ 0 0 35	63,892 7,382 0 0 3 2 3 2 90 109,038 46,730 0 0 3 2 3 2 90 47,943 32,322 0 0 3 2 3 2 90 49,291 17,307 0 0 3 2 3 2 90 46,037 12,1369 0 0 3 2 3 2 90 46,037 12,2477 0 0 3 2 3 2 90 32,619 14,480 0 0 3 2 3 2 90 56,786 16,893 0 0 3 2 3 2 90 72,102 17,681 0 0 3 2 3 2 90 79,639 12,501 0 0 3 2 3 2 90 79,639 12,501 0 0 3 3 3 3 3 3 3 </td <td>63,822 27,822 0 0 3 2 3 2 50 10 109,038 46,730 0 0 3 2 3 2 50 100 47,943 32,992 0 0 3 2 3 2 50 100 49,291 17,307 0 0 3 2 3 2 50 100 46,037 23,241 0 0 3 2 3 2 50 100 46,037 23,241 0 0 3 2 3 2 50 100 54,177 0 0 3 2 3 2 50 100 56,786 16,893 0 0 3 2 3 2 50 100 72,102 17,681 0 0 3 2 3 2 50 100 73,639 12,601 0 0 3 2 3 2 55 55 55 55 5</td> <td>63,89227,382003232501060109,03346,73000323250106047,94132,38200323250106047,94332,38200323250106069,51121,36900323250106064,01722,74100323250106054,21720,84000323250106054,21720,84000323250106054,21720,84000323250106054,21720,84000323250106072,1217,68100323250106072,1217,68100323250106072,1314833133455</td>	63,822 27,822 0 0 3 2 3 2 50 10 109,038 46,730 0 0 3 2 3 2 50 100 47,943 32,992 0 0 3 2 3 2 50 100 49,291 17,307 0 0 3 2 3 2 50 100 46,037 23,241 0 0 3 2 3 2 50 100 46,037 23,241 0 0 3 2 3 2 50 100 54,177 0 0 3 2 3 2 50 100 56,786 16,893 0 0 3 2 3 2 50 100 72,102 17,681 0 0 3 2 3 2 50 100 73,639 12,601 0 0 3 2 3 2 55 55 55 55 5	63,89227,382003232501060109,03346,73000323250106047,94132,38200323250106047,94332,38200323250106069,51121,36900323250106064,01722,74100323250106054,21720,84000323250106054,21720,84000323250106054,21720,84000323250106054,21720,84000323250106072,1217,68100323250106072,1217,68100323250106072,1314833133455

Russia (Pechora River)

Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (PECHORA RIVER). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

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Year Declared		Return	Return		Uncertainty in		Uncertainty in %	Estimated	Uncertainty in	Estimated	Uncertainty in
catch 1SW	catch MSW	estimate 1SW	estimate MSW	unreported	% unreported	unreported	unreported	exploitation	exploitation	exploitation	exploitation
salmor	salmon	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
				salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971 605	17,728	NA	NA	20	10	20	10	65	15	65	15
1972 825		NA	NA	20	10	20	10	65	15	65	15
1973 1,705		NA	NA	20	10	20	10	65	15	65	15
1974 1,320		NA	NA	20	10	20	10	65	15	65	15
1975 1,298		NA	NA	20	10	20	10	65	15	65	15
1976 991		NA	NA	20	10	20	10	65	15	65	15
1977 589	20,464	NA	NA	20	10	20	10	65	15	65	15
1978 759	26,341	NA	NA	20	10	20	10	65	15	65	15
1979 421	14,614	NA	NA	20	10	20	10	65	15	65	15
1980 1,123		NA	NA	20	10	20	10	65	15	65	15
1981 126		NA	NA	20	10	20	10	65	15	65	15
1982 54		NA	NA	20	10	20	10	65	15	65	15
1983 598		NA	NA	20	10	20	10	65	15	65	15
1984 1,833		NA	NA	20	10	20	10	65	15	65	15
1985 2,763	29,738	NA	NA	20	10	20	10	65	15	65	15
1986 66	32,734	NA	NA	20	10	20	10	65	15	65	15
1987 21	21,179	NA	NA	20	10	20	10	65	15	65	15
1988 3,184	12,816	NA	NA	20	10	20	10	65	15	65	15
1989 NA	NA	24,596	27,404	10	5	10	5	65	15	65	15
					5			65	15		15
	NA	50	49,950	10		10	5			65	
991 NA	NA	7,975	47,025	10	5	10	5	65	15	65	15
1992 NA	NA	550	54,450	10	5	10	5	65	15	65	15
1993 NA	NA	68	67,932	10	5	10	5	65	15	65	15
1994 NA	NA	3,900	48,100	10	5	10	5	65	15	65	15
1995 NA	NA	9,280	70,720	10	5	10	5	65	15	65	15
1996 NA	NA	8,664	48,336	10	5	10	5	65	15	65	15
1997 NA	NA	1,440	38,560	10	5	10	5	65	15	65	15
1998 NA	NA	780	59,220	10	5	10	5	65	15	65	15
1999 NA	NA	2,120	37,880	10	5	10	5	65	15	65	15
2000 NA	NA	84	83,916	10	5	10	5	65	15	65	15
2001 NA	NA	2,244	41,756	10	5	10	S	65,	15	65	15
2002 NA	NA	405	44,595	10	5	10	5	▲ 65	15	65	15
2003 NA	NA	1,650	31,350	10	5	10	5	65	15	65	15
2004 NA	NA	6,075	20,925	10	5	10		65	15	65	15
2005 NA		2,852	28,148	10	5	10		65	15	65	15
		1,472	30,528	10	5	10		65	15	65	15
2007 NA	NA	817	42,183	10	5	10	5	65	15	65	15
2008 NA	NA	300	49,700	10	5	10	5	65	15	65	15
2009 NA	NA	1,116	47,384	10	5	10	5	65	15	65	15
2010 NA	NA	1,096	53,704	10	5	10	5	65	15	65	15
2011 NA	NA	2,990	56,810	10	5	10	5	65	15	65	15
2012 NA		4,424	27,176	10	5	10	5	65	15	65	15
2013 NA		4,225	30,983	10		10	5	65	15	65	15
2013 NA 2014 NA		2,251	31,349	10			5	65	15		15
					~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~	10	5			65	
2015 NA		4,626	34,574	10	5	10	5	65	15	65	15
2016 NA	NA	4,260	31,840	10	5	10	5	65	15	65	1
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Sweden

Annual input data for NEAC PFA run-reconstruction & NCL models for SWEDEN. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

Uncertainty in exploitation Estimated exploitation Uncertainty in exploitation rate (%) - MSW rate (%) - MSW salmon 12.5 57.5 12.5	Estimated exploitation rate (%) - 15W 52.5 52.5 52.5 52.5 52.5 52.5 52.5 52	Uncertainty in % unreported catch of MSW salmon 15 15 15 15 15 15 15 15 15 15 15 15 15	Estimated % unreported catch of MSW salmon 30 30 30 30 30 30 30 30 30 30 30 30 30	Uncertainty in % unreported catch of 1SW salmon 15 15 15 15 15 15 15 15 15 15 15 15 15	Estimated % unreported catch of 15W 30 30 30 30 30 30 30 30 30 30 30 30 30	Declared catch MSW salmon 254 201 895 563 160 480 206 254 661 1,283	Declared catch 15W salmon 6,220 4,943 6,124 8,870 9,620 5,420 2,453 2,903 2,958 2,903 2,988 3,842 7,013	Year 1971 1972 1973 1974 1975 1976 1977 1978 1979 1980
rate (%) - 1SW rate (%) - MSW rate (%) - MSW salmon salmon salmon 12.5 \$7.5 12.5 12.5 \$7.5	rate (%) - 1SW salmon 52.5	catch of MSW salmon 15 15 15 15 15 15 15 15 15 15 15 15 15	catch of MSW salmon 30 30 30 30 30 30 30 30 30 30 30 30 30	catch of 1SW salmon 15 15 15 15 15 15 15 15 15 15 15 15 15	catch of 1SW salmon 30 30 30 30 30 30 30 30 30 30 30 30 30	salmon 254 201 895 563 160 480 206 254 661 1,283	6,220 4,943 6,124 8,870 9,620 5,420 2,453 2,903 2,988 3,842	1972 1973 1974 1975 1976 1977 1978 1979
salmon salmon salmon 12.5 57.5 12.5 12.5 50 15	salmon 52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.	salmon 15 15 15 15 15 15 15 15 15 15 15 15 15	salmon 30 30 30 30 30 30 30 30 30 30 30 30 30	salmon 15 15 15 15 15 15 15 15 15 15 15 15 15	salmon 30 30 30 30 30 30 30 30 30 30 30	254 201 895 563 160 480 206 254 661 1,283	6,220 4,943 6,124 8,870 9,620 5,420 2,453 2,903 2,988 3,842	1972 1973 1974 1975 1976 1977 1978 1979
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15 15 15 15 15 15 15 15 15 15 1	30 30 30 30 30 30 30 30 30 30 30 30 30 3	15 15 15 15 15 15 15 15 15 15 15 15 15	30 30 30 30 30 30 30 30 30 30	201 895 563 160 480 206 254 661 1,283	4,943 6,124 8,870 9,620 2,453 2,903 2,988 3,842	1972 1973 1974 1975 1976 1977 1978 1979
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15 15 15 15 15 15 15 15 15 15	30 30 30 30 30 30 30 30 30 30 30 30	15 15 15 15 15 15 15 15 15 15 15 15	30 30 30 30 30 30 30 30	895 563 160 480 206 254 661 1,283	4,943 6,124 8,870 9,620 2,453 2,903 2,988 3,842	1973 1974 1975 1976 1977 1978 1979
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15 15 15 15 15 15 15 15 15 15	30 30 30 30 30 30 30 30 30 30 30	15 15 15 15 15 15 15 15 15 15	30 30 30 30 30 30 30 30	563 160 480 206 254 661 1,283	8,870 9,620 5,420 2,453 2,903 2,988 3,842	1974 1975 1976 1977 1978 1979
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15 15 15 15 15 15 15 15 15	30 30 30 30 30 30 30 30 30	15 15 15 15 15 15 15 15 15	30 30 30 30 30 30	160 480 206 254 661 1,283	9,620 5,420 2,453 2,903 2,988 3,842	1975 1976 1977 1978 1979
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15 15 15 15 15 15 15	30 30 30 30 30 30 30 30	15 15 15 15 15 15 15 15	30 30 30 30 30	480 206 254 661 1,283	5,420 2,453 2,903 2,988 3,842	1976 1977 1978 1979
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15 15 15 15 15 15 15	30 30 30 30 30 30 30 30	15 15 15 15 15 15 15	30 30 30 30	206 254 661 1,283	2,453 2,903 2,988 3,842	1977 1978 1979
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15 15 15 15 15	30 30 30 30 30 30	15 15 15 15 15	30 30 30	254 661 1,283	2,903 2,988 3,842	1978 1979
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15 15 15 15	30 30 30 30 30	15 15 15 15	30 30	661 1,283	2,988 3,842	1979
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15 15 15	30 30 30 30	15 15 15	30	1,283	3,842	
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15 15	30 30 30	15 15				1980
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15 15	30 30	15	30			
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5 52.5	15 15 15	30			284		1981
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5 52.5 52.5	15 15			30	1,381	6,177	1982
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5 52.5	15	30	15	30	903	8,222	1983
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5 52.5		20	15	30	1,266	11,584	1984
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	52.5 52.5	15	30	15	30	470	13,810	1985
12.5 57.5 12.5 12.5 57.5 12.5 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 12.5 42.5 12.5	52.5	15	30	15	30	240	14,415	1986
12.5 57.5 12.5 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 42.5 12.5		15 15	30 30	15 15	30 30	1,084 1,160	11,450 9,604	1987 1988
15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 15 50 15 12.5 42.5 12.5		15	30	15	30	4,044	2,803	1989
15 50 15 15 50 15 15 50 15 15 50 15 12.5 42.5 12.5	45	10	15	10	15	2,249	6,839	1990
15 50 15 15 50 15 15 50 15 15 50 15 12.5 42.5 12.5	45	10	15	10	15	3,033	8,599	1991
15 50 15 15 50 15 12.5 42.5 12.5	45	10	15	10	15	4,205	9,550	1992
15 50 15 12.5 42.5 12.5	45	10	15	10	15	4,762	9,468	1993
12.5 42.5 12.5	45	10	15	10	15	3,628	7,347	1994
12.5 42.5 12.5	37.5	10	15	10	15	1,528	8,933	1995
	37.5	10	15	10	15	2,507	5,318	1996
12.5 42.5 12.5	37.5	10	15	10	15	1,809	2,415	1997
12.5 42.5 12.5	37.5	10 -	15	10	15	1,000	1,953	1998
12.5 42.5 12.5	37.5	10	15	10	15	712	3,075	1999
12.5 42.5 12.5	37.5	10	15	10	15	2,546	5,660	2000
12.5 42.5 12.5	37.5	10	15	10	15	3,026	3,504	2001
12.5 42.5 12.5	37.5	10	15	10	15	2,075	3,374	2002
12.5 42.5 12.5	37.5	10	15	10	15	496	1,833	2003
12.5 42.5 12.5	37.5	10	15	10	15	1,528	1,537	2004
12.5 42.5 12.5	37.5	10	15	10	15	1,027	1,503	2005
12.5 42.5 12.5	37.5	10	15 15	10	15	1,069	1,676	2006 2007
12.5 42.5 12.5 12.5 32.5 12.5	37.5 27.5	10 7.5	12.5	10	15 12.5	1,001 1,112	521 615	2007
12.5 32.5 12.5 12.5 32.5 12.5	27.5	7.5	12.5	7.5 7.5	12.5	979	615	2008
12.5 32.5 12.5 12.5 32.5 12.5	27.5	7.5	12.5	7.5	12.5	1,139	1,111	2009
15 50 15	45	7.5	12.5	7.5	12.5	3,100	1,111	2010
12.5 32.5 12.5	27.5	7.5	17.5	7.5	12.5	3,130	1,460	2011
15 35 15	30	5	10	5	10	1,431	874	2012
12.5 35 12.5	30	7.5	12.5	7.5	12.5	2,797	2,347	2013
12.5 30 12.5	30	7.5	12.5	.5	12.5	2,569	1,028	2015
10 25 10	25	7.5	10	7.5	10	910	554	2016

ICES Stock Annex UK (England and Wales)

Annual input data for NEAC PFA run-reconstruction & NCL models for UK(ENGLAND AND WALES). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Estimated	Declared catch	Declared catch	Declared catch	Estimated	Estimated % l	Incertainty in %	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in	Estimated %	Estimated	Estimated	Estimated
	total catch	proportion	in NE coastal	in NE coastal	in NE coastal		unreported	unreported	unreported	% unreported	exploitation	exploitation	•	xploitation rate	unreported	proportion	proportion	proportion
		1SW (total)	fishery -	fishery -	fishery -	1SW (NE	catch of 1SW	catch of 1SW	catch of MSW		rate (%) - 1SW	. ,	rate (%) - MSW	(%) - MSW		Scottish fish in		
			total	drift nets	T/J nets	fishery)	salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon	fishery	NE fishery	NE fishery	
1071	400.004	0.55	60.050			0.55	20.2		20.2		57.0	10	12.5	10		(total)	(drift)	(T/J nets)
1971	109,861	0.55	60,353	NA	NA	0.55	38.3	9.6	38.3	9.6	57.3	10	42.5	10	32.3	0.95	NA	NA
1972 1973	108,074	0.42 0.53	51,681 62,842	NA	NA	0.42 0.53	39 38.4	9.7	39 38.4	9.7	51.3 50.6	10	37.8 37.3	10 10	32.3	0.95	NA	NA
1973	114,786 104,325	0.53	62,842 52,756	NA NA	NA NA	0.53	38.4 39.3	9.6 9.8	38.4 39.3	9.6 9.8	50.6	10 10	37.3	10	32.3 32.3	0.95 0.95	NA NA	NA NA
1974	104,323	0.65	52,750	NA	NA	0.65	38.5	9.8	39.5	9.6	10.2	10	36.7	10	32.3	0.95	NA	NA
1975	54,294	0.59	15,701	NA	NA	0.59	36.8	9.2	36.8	9.2	50.3	10	37.1	10	32.3	0.93	NA	NA
1970	94,294 94,282	0.62	52,888	NA	NA	0.62	39	9.8	39	9.8	50.3	10	37.1	10	32.3	0.94	NA	NA
1978	93,125	0.69	51,630	NA	NA	0.69	38.4	9.6	38.4	9.6	40.1	10	36.2	10	32.3	0.92	NA	NA
1979	75,386	0.81	43,464	NA	NA	0.81	38.6	9.6	38.6	9.6	47.7	10	35.2	10	32.3	0.92	NA	NA
1980	90,218	0.55	45,780	NA	NA	0.55	39.1	9.8	39.1	9.8	47.8	10	35.2	10	32.3	0.90	NA	NA
1981	121,039	0.48	69,113	NA	NA	0.48	38.3	9.6	38.3	96	47.4	10	34.9	10	32.3	0.89	NA	NA
1982	80,289	0.67	50,167	NA	NA	0.67	38.3	9.6	38.3	96	473	10	34.8	10	32.3	0.88	NA	NA
1983	116,995	0.72	77,277	NA	NA	0.72	37.1	9.3	37.1	9.3	47.1	10	34.7	10	32.3	0.87	NA	NA
1984	94,271	0.74	59,295	NA	NA	0.74	36.5	9.1	36.5	9.1	47.4	10	34.8	10	32.3	0.86	NA	NA
1985	95,531	0.66	57,356	NA	NA	0.66	38.9	9.7	38.9	9.7	47.5	10	34.9	10	32.3	0.85	NA	NA
1986	110,794	0.62	63,425	NA	NA	0.62	38	9.5	38	9.5	46.9	10	34.3	10	32.3	0.84	NA	NA
1987	83,439	0.68	36,143	NA	NA	0.68	38.2	9.5	38.2	9.5	46.1	10	33.7	10	32.3	0.83	NA	NA
1988	110,163	0.69	NA	47,465	3,384	0.69	39.7	9.9	39.7	9.9	45.5	10	33.5	10	32.3	NA	0.82	0.50
1989	83,668	0.65	NA	36,236	5,217	0.65	36.9	9.2	36.9	9.2	45.3	10	33.3	10	32.3	NA	0.81	0.50
1990	86,676	0.52	NA	48,219	3,311	0.52	36.7	9.2	36.7	9.2	45.3	10	33.2	10	31.3	NA	0.80	0.50
1991	51,649	0.71	NA	22,463	2,966	0.71	37.3	9.3	37.3	9.3	44.0	10	32.3	10	29.7	NA	0.79	0.50
1992	44,586	0.77	NA	17,574	2,570	0.77	39.8	10	39.8	10	43.5	10	31.8	10	28	NA	0.78	0.50
1993	69,177	0.81	NA	39,224	2,576	0.81	38	9.5	38	9.5	40.6	10	29.5	10	26.3	NA	0.77	0.50
1994	88,121	0.77	NA	41,298	5,256	0.77	23.9	6.0	23.9	6	40.5	10	29.5	10	24.4	NA	0.76	0.50
1995	80,478	0.72	NA	48,005	5,205	0.72	22.3	5.6	22.3	5.6	37.6	10	27.1	10	22.5	NA	0.75	0.50
1996	46,696	0.65	NA	15,172	3,409	0.65	20.6	5.1	20.6	5.1	35.8	10	25.8	10	20.6	NA	0.75	0.50
1997	41,374	0.73	NA	19,241	2,681	0.73	18.8	4.7	18.8	4.7	33.4	10	23.9	10	18.5	NA	0.75	0.50
1998	36,917	0.82	NA	17,328	937	0.82	18.9	4.7	18.9	4.7	31.4	10	22.4	10	18.5	NA	0.75	0.50
1999	41,094	0.68	NA	24,812	2,021	0.68	17.4	4.4	17.4	4.4	29.5	10	17.9	9	17.1	NA	0.75	0.50
2000	60,953	0.79	NA	40,059	3,295	0.79	14.9	3.7	14.9	3.7	29.7	10	15	7.5	13.1	NA	0.75	0.50
2001	51,307	0.75	NA	32,374	3,741	0.75	14.8	3.7	14.8	3.7	27.9	10	14.3	7.1	13.1	NA	0.75	0.50
2002	45,669	0.76	NA	27,685	3,295	0.76	15.3	3.8	15.3	3.8	27.8	10	14.1	7	13.9	NA	0.75	0.50
2003	22,206	0.66	NA	5,511	4,924	0.66	17.4	4.4	17.4	4.4	21.4	10	10.7	5.3	17.1	NA	0.75	0.50
2004	30,559	0.81	NA	5,921	5,096	0.81	17.7	4.4	17.7	4.4	22.1	10	10.6	5.3	17.1	NA	0.75	0.50
2005	26,162	0.76	NA	5,607	3,380	0.76	17.6	4.4	17.6	4.4	21.8	10	10.6	5.3	17.1	NA	0.75	0.50
2006	22,056	0.78	NA	4,040	3,526	0.78	17.6	4.4	17.6	4.4	19.5	9.8	9.1	4.6	17.1	NA	0.75	0.50
2007	19,914	0.78	NA	4,894	2,197	0.78	17.7	4.4	17.7	4.4	17.9	9	8.4	4.2	17.1	NA	0.75	0.50
2008	19,036	0.76	NA	3,649	2,592	0.76	17.8	4.4	17.8	4.4	17.6	8.8	8.2	4.1	17.1	NA	0.75	0.50
2009	13,910	0.72	NA	2,590	2,805	0.72	11.4	2.9	11.4	2.9	17.4	8.7	8.2	4.1	7.4	NA	0.75	0.50
2010	32,695	0.78	NA	12,214	7,768	0.78	10.8	2.7	10.8	2.7	17.5	8.8	8	4	7.4	NA	0.75	0.50
2011	34,575	0.57	NA	14,915	9,233	0.57	10.1	2.5	10.1	2.5	20.8	10	10.2	5.1	7.4	NA	0.64	0.37
2012	14,926	0.50	NA	3,571	3,705	0.50	11.2	2.8	11.2	2.8	16.8	8.4	8	4	7.4	NA	0.62	0.37
2013	22,608	0.58	NA	7,964	8,679	0.58	9.5	2.4	9.5	2.4	17.4	8.7	8.5	4.3	7.4	NA	0.63	0.37
2014	14,219	0.54	NA	6,974	3,826	0.54	9.3	2.3	9.3	2.3	15.8	7.9	8	4	7.4	NA	0.64	0.37
2015	19,262	0.47	NA	9,233	6,657	0.47	12.9	3.2	12.9	3.2	15.2	7.6	7.7	3.9	7.4	NA	0.64	0.37
2016	22,471	0.41	NA	10,811	7,956	0.41	11.9	3	11.9	3	14.8	7.4	7.5	3.7	7.4	NA	0.63	0.37

UK (Northern Ireland)-Foyle Fisheries Area

Annual input data for NEAC PFA run-reconstruction & NCL models for UK(NORTHERN IRELAND) (FOYLE). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Declared	Declared rod	Declared rod		Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	
	net catch	net catch	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation		exploitatio
	1SW salmon	MSW salmon	salmon	salmon	catch of 1SW salmon	catch of 1SW salmon	catch of MSW salmon	catch of MSW salmon	rate (%) - 1SW salmon	salmon	rate (%) - MSW salmon	salmo
1971	78,037	5,874	NA	NA	21.5	11.5	21.5	11.5	80	5		
1972	64,663	4,867	NA	NA	21.5	11.5	21.5	11.5	80	5		
1973	57,469	4,326	NA	NA	21.5	11.5	21.5	11.5	80	5		
1974	72,587	5,464	NA	NA	21.5	11.5	21.5	11.5	80	5		
1975	51,061	3,843	NA NA	NA NA	21.5	11.5	21.5	11.5	80 80	5		
1976 1977	36,206 36,510	2,725 2,748	NA	NA	21.5 21.5	11.5 11.5	21.5 21.5	11.5 11.5	80	5		
1978	44,557	3,354	NA	NA	21.5	11.5	21.5	11.5	80	5		
1979	34,413	2,590	NA	NA	21.5	11.5	21.5	11.5	80	5		
1980	45,777	3,446	NA	NA	21.5	11.5	21.5	11.5	80	5	50	
1981	32,346	2,435	NA	NA	21.5	11.5	21.5	11.5	80	5		
1982	55,946	4,211	NA	NA	21.5	11.5	21.5	11.5	80	5		
1983	77,424	5,828	NA	NA	21.5	11.5	21.5	11.5	80	5		
1984 1985	27,465 37,685	2,067 2,836	NA NA	NA NA	21.5 21.5	11.5 11.5	21.5 21.5	11.5 11.5	80 80	5 5		
1985	43,109	3,245	NA	NA	21.5	11.5	21.5	11.5	80	5		
1987	17,189	1,294	NA	NA	21.5	11.5	21.5	11.5	69	7	46	
1988	43,974	3,310	NA	NA	21.5	11.5	21.5	11.5	65	7	36	
1989	60,288	4,538	NA	NA	23.5	13.5	23.5	13.5	89	9	60	
1990	39,875	3,001	NA	NA	13.5	3.5	13.5	3.5	62	6		
1991	21,709	1,634	NA	NA	13.5	3.5	13.5	3.5	65	7	43	
1992 1993	39,299 35,366	2,958 2,662	NA NA	NA NA	16.5 13.5	6.5 3.5	16.5 13.5	6.5 3.5	56 41	6 4		
1993	36,144	2,002	NA	NA	13.5	3.5	13.5	3.5	70	7	40	
1995	33,398	2,514	NA	NA	13.5	3.5	13.5	3.5	67	7	42	
1996	28,406	2,138	NA	NA	15	5	15.0	5	57	10		
1997	40,886	3,077	NA	NA	10	5	10	5	60	10	34	
1998	37,154	2,797	NA	NA	10	5	10	5	25	5		1
1999	21,660	1,630	NA	NA	10	5	10	5	63	5		-
2000 2001	30,385 21,368	2,287 1,608	NA NA	NA NA	10 5	5 5	10	55	58 50	5		7
2001	37,914	2,854	9,163	690	2.5	2.5	2,5	2.5	15	5		
2003	30,441	2,291	4,576	344	0.5	0.5	0.5	0.5	15	3		
2004	20,730	1,560	4,570	344	0.5	0.5	0.5	0.5	15	3		
2005	23,746	1,787	7,079	533	0.5	0.5	0.5	0.5	15	3		
2006	11,324	852	4,886	368	0.5	0.5	0.5	0.5	15	3		
2007	5,050	322	9,530	608	0.5	0.5	0.5	0.5	15	3		
2008 2009	3,880 1,743	292 194	4,755 3,640	304 405	0.5 0.5	0.5	0.5	0.5 0.5	15 15	3		
2009	1,743	0	3,040	388	0.5	0.5	0.5	0.5	15	3		
2011	0	0	2,276	759	1			1	15	5		
2012	0	0	4,781	1,594	1	1	1	1	10	7.5		
2013	0	0	5,030	498	1	1	1	1	10	7.5		1
2014	0	0	2,029	225	1	1	1	1	10	7.5		7
2015	0	0	1,998	250	1	1	1	1	10	7.5		-
2016	0	0	3,457	432		1	1	1	10	7.5	10	

UK (Northern Ireland)-DAERA area.

Annual input data for NEAC PFA run-reconstruction & NCL models for UK(NORTHERN IRELAND) (DAERA Region) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Declared	Declared rod	Declared rod	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in C	ount River	Estimated	Count	Estimated	Count	Uncertainty	Count	Uncertai
icui	net catch	net catch	catch 1SW	catch MSW		% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation	Bush		River Bann			in upscaling		
	1SW salmon	MSW	salmon	salmon			catch of MSW							(%) 1SW in		(%) 1SW in		factor (%)	factor	factor (
		salmon			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon		Bush		Bann	(%) 1SW	1SW	(%) MSW	
971	35,506	2,673	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5) () () 0) NA	NA NA	NA	
.972	34,550	2,601	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5) () (0) NA	NA NA	NA	
.973	29,229	2,200	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	C) C	0) NA	NA NA	NA	
1974	22,307	1,679	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	C		D C	0) NA	NA NA	NA	
1975	26,701	2,010	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	9) () 0) NA	NA NA	NA	
1976	17,886	1,346	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5		Y (0 0) 0) NA	NA NA	NA	
1977	16,778	1,263	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5) (0 0) 0) NA	NA NA	NA	
1978	24,857	1,871	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	🔺 _ C) () () 0) NA	NA NA	NA	
1979	14,323	1,078	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	0) (0 0) 0) NA	NA NA	NA	
1980	15,967	1,202	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	C) () () 0) NA	NA NA	NA	
1981	15,994	1,204	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	0) (0 0) 0) NA	NA NA	NA	
1982	14,068	1,059	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	0) (0 0) 0) NA	NA NA	NA	
1983	20,845	1,569	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	0) (0 0) 0) NA	NA NA	NA	
1984	11,109	836	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	() () () 0) NA	NA NA	NA	
1985	12,369	931	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	0) (0 0) 0) NA	NA NA	NA	
986	13,160	991	NA	NA	21.5	11.5	21.5	11.5	80	5	50	5	() () () () NA	NA NA	NA	
987	9,240	695	NA	NA	21.5	11.5	21.5	11.5	69	7	46	5	2,530) 83	з с) 0) 10) 2	9.5	
988	14,320	1,078	NA	NA	21.5	11.5	21.5	11.5	64.5	6.5	36	4	2,832	2 96	6 C) () 10) 2	9.5	
989	15,081	1,135	NA	NA	23.5	13.5	23.5	13.5	. 89	9	60	6	1,029	82	2 0) 0) 10) 2	9.5	
990	9,499	715	NA	NA	13.5	3.5	13.5	3.5	62	6	38	4	1,850) 87	7 C	0) 10) 2	9.5	
.991	6,987	526	NA	NA	13.5	3.5	13.5	3.5	64.5	6.5	43	4	2,341	87	7 C	0) 10) 2	9.5	
.992	9,346	703	NA	NA	16.5	6.5	16.5	6.5	56	6	33	3	2,546	5 84	4 C	0) 10) 2	9.5	
1993	7,906	595	NA	NA	13.5	3.5	13.5	3.5	41	4	12	1	3,235	93	з с	0) 10) 2	9.5	
1994	11,206	843	NA	NA	19	9	19	9	70	7	40	4	2,010) 88	в с	0) 10) 2	9.5	
1995	11,637	876	NA	NA	13.5	3.5	13.5	3.5	67	7	42	4	1,521			0) 10) 2	9.5	
1996	10,383	781	NA	NA	15	5	15	5	57	10	34	10	1,097	87	7 C) 0) 10) 2	9.5	
1997	10,479	789	NA	NA	10	5	10	5	60	10	34	10	1,677						61	
1998	9,375	706	NA	NA	10	5	10	5	25	5	22.5	7.5	2,995	95	5 11,462	95	67	5	61	
.999	9,011	678	NA	NA	10	5	10	5	63	5	32.5	7.5	977) 67	5	61	
000	10,598	798	NA	NA	10	5	10	5	58	5	32.5	7.5	950) 91	1 5,979	91	67	5	61	
001	8,104	610	NA	NA	5	5	5	5	50	5	30	5	913	91	7 5,771	. 97	67	5	61	
002	3,315	249	2,218	167	2.5	2.5	2.5	2.5	13.7	8.8	13.7	8.8	835			95	67	5	61	
003	2,236	168	1,884	141	2.5	2.5	2.5	2.5	12.3	6.6	12.3	6.6	723			96	6 67	5	61	
004	2,411	181	3,053	230	0.5	0.5	0.5	0.5	18.3	9.7	18.3	9.7	878	92	2 9,050	92	2 67	5	61	
005	3,012	227	1,791	135	0.5	0.5	0.5	0.5	11.9	7.1	11.9	7.1	1,151	91	1 6,609	91	67	5	61	
006	2,288	172	1,289	97	0.5	0.5	0.5	0.5	12.4	8	12.4	8	1,074			87	67	5	61	
2007	2,533	162	2,427	155	0.5	0.5	0.5	0.5	11	3.6	11	3.6	2,584						61	
800	1,825	116	2,444	156	0.5	0.5	0.5	0.5	13.9	7.1	13.9	7.1	1,712			-			9.5	
009	1,383	154	1,457	162	0.5	0.5	0.5	0.5	9.9	3	9.9	3	726						61	
2010	1,723	191	1,327	147	0.5	0.5	0.5	0.5	14.6	2.5	14.6	2.5	1,045						61	
011	857	285	1,132	378	1	1	1	1	15	5	15	5	649						61	
012	15	5	263	87	1	1	1	1	10	5	10	5	926				67	5	61	
013	9	1	46	5	1	1	1	1	0.6	0.4	0.6	0.4	1,644	92	2 5,866	91			61	
2014	0	0	143	40	2.5	2.5	2.5	2.5	0.6	0.4	0.6	0.4	963						61	
2015	0	0	20	6	2.5	2.5	2.5	2.5	0.6	0.4	0.6	0.4	1,005						61	
016	0	0	124	279	2.5	2.5	2.5	2.5	0.6	0.4	0.6	0.4	2,015	8	5 15,936	86	67	, 5	61	

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UK (Scotland)-East

Annual input data for NEAC PFA run-reconstruction & NCL models for SCOTLAND (EAST) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

I Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in	Estimated %	Uncertainty in	Estimated %	Declared	Declared	Year
	exploitation	exploitation	exploitation	% unreported	unreported	% unreported	unreported	catch MSW	catch 1SW	
	rate (%) - MSW		rate (%) - 1SW	catch of MSW	catch of MSW	catch of 1SW	catch of 1SW	salmon	salmon	
. ,	salmon	salmon	salmon	salmon	salmon	salmon	salmon			
) 1	49.9	12.6	75.4	10	25	10	25	135,530	216,873	1971
5 10.	51.5	12.8	76.8	10	25	10	25	183,875	220,106	1972
) 1	49.9	12.5	74.9	10	25	10	25	204,826	259,773	1973
11.	56.4	13.7	82.0	10	25	10	25	158,959	245,424	1974
L 1	55.1	13.4	80.5	10	25	10	25	180,828	181,940	1975
10.	50.7	12.8	76.6	10	25	10	25	92,179	150,069	1976
3 11.	55.8	13.6	81.4	10	25	10	25	118,645	154,306	1977
10.	51	12.6	75.6	10	25	10	25	139,763	158,859	1978
9 10.	53.9	13.1	78.4	10	25	10	25	116,559	160,796	1979
2 10	52	12.8	76.8	7.5	17.5	7.5	17.5	155,646	101,665	1980
10.	51.2	12.7	75.9	7.5	17.5	7.5	17.5	156,683	129,690	1981
8 9.	45.3	11.8	71.1	7.5	17.5	7.5	17.5	113,198	175,374	1982
9.	49.4	12.8	77.0	7.5	17.5	7.5	17.5	126,104	170,843	1983
8.	43.9	11.7	70.1	7.5	17.5	7.5	17.5	90,829	175,675	1984
7.	38.9	10.3	61.9	7.5	17.5	7.5	17.5	95,044	133,119	1985
57.	37.6	9.9	59.5	7.5	17.5	7.5	17.5	128,654	180,292	1986
5 8.	40.5	10.8	64.5	7.5	17.5	7.5	17.5	88,531	139,259	1987
2 5.	29.2	6.7	40.3	7.5	17.5	7.5	17.5	91,167	118,620	1988
3 5.	28	6.3	37.5	5	10	5	10	85,399	143,063	1989
7 5.	28.7	6.6	39.8	5	10	5	10	73,974	63,352	1990
5 5.	27.5	6.1	36.8	5	10	5	10	53,693	53,861	1991
5.	25.9	5.4	32.1	5	10	5	10	67,979	79,906	1992
) 5	26.9	5.9	35.3	5	10	5	10	60,498	73,413	1993
L 5.	26.1	5.5	33.1	5	10	5	10	72,758	80,429	1994
1 5.	25.4	5.2	30.9	5	10	5	10	69,055	73,020	1995
5 4.	24.5	4.8	28.8	5	10	5	10	50,365	56,627	1996
L I	25.1	5.1	30.6	51	10	5	10	34,850	37,448	1997
9 4.	22.9	4	24.2	5	10	5	10	32,231	44,952	1998
3 4.	23.3	4.1	24.8	5	10	5	10	27,011	20,907	1999
3 4.	22.3	3.6	21.9	5	10	5	10	31,280	36,871	2000
5 4.	21.6	3.4	20.4	5	10	5	10	30,470	36,646	2001
2 4.	21.2	3.2	19.3	5	10	5	10	21,740	26,616	2002
3 4.	19.3	2.8	17.3	5	10	5	10	24,270	25,871	2003
3 4.	19.3	2.8	17.3	5	10	5	10	30,773	31,667	2004
3 4.	19.3	2.8	17.3	5	10	5	10	23,676	31,597	2005
3.	16.5	2.8	15.3	5	10	5	10	22,954	30,739	2006
5 3.	15	2.8	13.8	5	10	- 5	10	19,444	26,015	2007
J 3.	14	2.8	10.8	5	10	5	10	20,757	18,586	2008
3.	13	2.8	9.8	5	10	5	10	15,042	14,863	2009
3.	13	2.8	9.8	5	10	5	10	22,861	28,161	2010
5 3.	12.5	2.8	9.3	5	10	5	10	24,218	12,495	2011
5 3.	11.5	2.8	8.3	5	10	5	10	16,165	16,117	2012
3.	11	2.8	7.3	5	10	5	10	14,901	18,400	2013
5	9.5	2.3	5.8	5	10	5	10	11,798	10,924	2014
5	9.5	2.3	5.8	5	10	5	10	7,789	12,223	2015
5	5	0.5	2	5	10	5	10	4,013	4,015	2016

UK (Scotland)-West

Annual input data for NEAC PFA run-reconstruction & NCL models for SCOTLAND (WEST) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

Uncertainty exploitati	Estimated exploitation	Uncertainty in exploitation	Estimated exploitation	Uncertainty in % unreported	Estimated % unreported	Uncertainty in % unreported	Estimated % unreported	Declared catch MSW	Declared catch 1SW	Year
	rate (%) - MSW		rate (%) - 1SW	catch of MSW	catch of MSW	catch of 1SW	catch of 1SW	salmon	salmon	
								Salilion	Saimon	
salm	salmon 24.9	salmon 6.3	salmon 37.7	salmon 10	salmon 35	salmon 10	salmon 35	26,071	45,287	971
	24.5	6.4	38.4	10	35	10	35	34,148	31,358	.972
	23.7	6.2	37.5	10	35	10	35	33,094	33,317	973
	24.9	6.8	41	10	35	10	35	29,369	43,992	973 974
			41 40.2		35		35			974 975
	27.5	6.7		10		10		27,060	40,217	
	25.3	6.4	38.3	10	35	10	35	22,292	38,254	.976
	27.9	6.8	40.7	10	35	10	35	20,286	39,739	977
	25.5	6.3	37.8	10	35	10	35	23,085	45,342	.978
	26.9	6.5	39.2	10	35	10	35	15,923	26,284	979
	26	6.4	38.4	7.5	27.5	7.5	27.5	16,882	19,654	980
	25.6	6.3	38	7.5	27.5	7.5	27.5	17,896	20,642	981
	22.6	5.9	35.5	7.5	27.5	7.5	27.5	15,005	32,352	982
	24.7	6.4	38.5	7.5	27.5	7.5	27.5	19,778	38,406	.983
	21.9	5.8	35.1	7.5	27.5	7.5	27.5	16,321	37,262	984
	19.5	5.2	30.9	7.5	27.5	7.5	27.5	19,521	24,648	985
	18.8	5	29.8	7.5	27.5	7.5	27.5	19,507	22,483	986
	20.3	5.4	32.3	7.5	27.5	7.5	27.5	15,446	25,388	987
	14.6	3.4	20.2	7.5	27.5	7.5	27.5	20,972	30,312	988
	14	3.1	18.8	5	20	5	20	18,474	31,788	989
	14.4	3.3	19.9	5	20	5	20	13,942	17,729	.990
	13.7	3.1	18.4	5	20	5	20	11,500	19,727	991
	12.9	2.7	16.1	5	20	5	20	14,862	21,770	992
	13.5	2.9	<u>1</u> 7.7	5	20	5	20	11,228	21,104	993
	13	2.8	16.5	5	20	5	20	12,304	18,234	994
	12.7	2.6	15.5	5.	20	5	20	9,133	16,784	995
	12.2	2.4	14.4	5	20	5	20	7,463	9,537	.996
	12.6	2.6	15.3	5	20	5	20	5,504	9,059	997
	11.5	2.0	12.1	5	20	5	20	6,150	8,369	998
	11.5	2.1	12.4	5	20	5	20	3,588	4,147	999
	11.7	1.8	10.9	5	20	5	20	5,301	6,974	000
	10.8	1.8	10.9	5	20	5	20	4,191	5,603	000
		1.7	9.6			5	20			
	10.6			5	20			4,548	4,691	002
	5.3	0.8	4.8	5	20	5	20	3,061	3,536	003
	7.5	1	7	5	20	5	20	6,024	5,836	004
	7.5	1	7	5	20	5	20	4,913	7,428	005
	7.5	1	7	5	20	5	20	4,403	5,767	006
	7.5	1	7	5	20	5	20	4,468	6,178	007
	7.5	1	7	5	20	5	20	4,853	4,740	008
	6.5	1	6	5	20	5	20	4,095	3,250	009
	6.5	1	6	5	20	5	20	4,051	5,107	010
	6	1	5.5	5	20	5	20	4,246	3,206	011
	5	1	4.5	5	20	5	20	3,392	3,239	012
	4.8	1	4	5	20	5	20	2,286	2,342	013
	4.5	1	3.5	5	20	5	20	1,519	1,653	014
	4.5	1	3.5	5	20	5	20	1,335	1,429	015
	0.5	0.3	0.7	5	20	5	20	137	278	016

Faroes

Annual input data for NEAC PFA run-reconstruction & NCL models for Faroes. (Uncertainty values define uniform distribution around estimate used in Monte Carlo simulation)

Year	Catch	Catch	Estimated %	Uncertainty in	% wild	Natural
	1SW salmon	MSW salmon	unreported	, % unreported		mortality after
			catch of 1SW	catch_of_1SW		, 1st sea winter
			salmon	salmon		(M)
1971	2,620	105,796	10	5	1	0.03
1972	2,754	111,187	10	5	1	0.03
1973	3,121	126,012	10	5	1	0.03
1974	2,186	88,276	10	5	1	0.03
1975	2,798	112,984	10	5	1	0.03
1976	1,830	73,900	10	5	1	0.03
1977	1,291	52,112	10	5	1	0.03
1978	974	39,309	10	5	1	0.03
1979	1,736	70,082	10	5	1	0.03
1980	4,523	182,616	10	5	1	0.03
1981	7,443	300,542	10	5	1	0.03
1982	6,859	276,957	10	5	1	0.03
1983	15,861	215,349	10	5	1	0.03
1984	5,534	138,227	10	5	1	0.03
1985	378	158,103	10	5	0.9	0.03
1986	1,979	180,934	10	5	1	0.03
1987	90	166,244	10	5	1	0.03
1988	8,637	87,629	10	5	0.9	0.03
1989	1,788	121,965	10	5	0.8	0.03
1990	1,989	140,054	10	5	0.5	0.03
1991	943	84,935	10	5	0.5	0.03
1992	68	35,700	10	5	0.6	0.03
1993	6	30,023	10	5	0.7	0.03
1994	15	31,672	10	5	0.7	0.03
1995	18	34,662	10	5	0.8	0.03
1996	101	28,381	10	5	0.8	0.03
1997	0	0	0	0	0	0.03
1998	339	1,424	15	5	0.8	0.03
1999	0	0	0	0	0	0.03
2000	225	1,765	15	5	0.8	0.03
2001	0	0	0	0	0	0.03
2002	0	0	0	0	0	0.03
2003	0	0	0	0	0	0.03
2004	0	0	0	0	0	0.03
2005	0	0	0	0	0	0.03
2006	0	0	0	0	0	0.03
2007	0	0	0	0	0	0.03
2008	0	0	0	0	0	0.03
2009	0	0	0	0	0	0.03
2010	0	0	0	0	0	0.03
2011	0	0	0	0	0	0.03
2012	0	0	0	0	0	0.03
2013		0	0	0	0	0.03
2014		0	0	0	0	0.03
2015	0	0	0	0	0	0.03
2016	0	0	0	0	0	0.03

West Greenland

Annual input data for NEAC PFA run-reconstruction & NCL models for WEST GREENLAND. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

	Declared catch (t)	Estimated unreported	wean weight	Estimated min' proportion of	Estimated max' proportion of	Proportion 1SW in NAC	Proportion 1SW in NEAC	No. Fish identified as NAC (from	No. Fish idstoffed composition as NEAC (from	
		catch (t)		NAC fish (from	NAC fish (from	fish	fish	genetic analysis)	as NEAC form Country genetic analysis)	MSW
				scale analysis)	scale analysis)				France	0.02
1971	2,689	0	3.14	0.28	0.4	0.945	0.964	NA	Finland	0.00
1972	2,113	0	3.44	0.34	0.37	0.945	0.964	NA	Iceland	0.00
1973 1974	2,341 1,917	0	4.18 3.58	0.39 0.39	0.59 0.46	0.945 0.945	0.964 0.964	NA	Irelano,d	0.14
1974	2,030	0	3.12	0.35	0.40	0.945	0.964	NA		0.02
1976	1,175	0	3.04	0.38	0.48	0.945	0.964	NA	Nor₩ay	
1977	1,420	0	3.21	0.38	0.57	0.945	0.964	NA	Russia	0.00
1978	984	0	3.35	0.47	0.57	0.945	0.964	NA	Sweden	0.0
1979	1,395	0	3.34	0.48	0.52	0.945	0.964	NA	UK(È₩gland & Wales)	0.1
1980	1,194	0	3.22	0.45	0.51	0.945	0.964	NA	UK(Northern Ireland)	0.0
1981 1982	1,264 1,077	0	3.17	0.58	0.61	0.945	0.964 0.964	NA	UK(Scotland)	0.6
1982	310	0	3.11 3.10	0.6 0.38	0.64 0.41	0.945 0.945	0.964	NA	NA	0.0
1984	297	0	3.11	0.47	0.53	0.945	0.964	NA		
1985	864	0	2.87	0.46	0.53	0.925	0.95	NA	Other	
1986	960	0	3.03	0.48	0.66	0.951	0.975	NA	NA	
1987	966	0	3.16	0.54	0.63	0.963	0.98	NA	TotallA	1.0
1988	893	0	3.18	0.38	0.49	0.967	0.981	NA	NA	
1989	337	0	2.87	0.52	0.6	0.923	0.955	NA	NA	
1990	274	0	2.69	0.7	0.79	0.957	0.963	NA	NA	
1991 1992	472 237	0	2.65 2.81	0.61 0.5	0.69 0.57	0.956 0.919	0.934 0.975	NA NA	NA NA	
1992	257	12	2.81	0.5	0.57	0.919	0.975	NA	NA	
1994	0	12	2.73	0.5	0.76	0.95		NA NA	NA	
1995	83	20	2.56	0.65	0.72	0.968	0.973	NA	NA	
1996	92	20	2.88	0.71	0.76	0.941	0.961	NA	NA	
1997	58	5	2.71	0.75	0.84	0.982	0.993	NA	NA	
1998	11	11	2.78	0.73	0.84	0.968	0.994		NA	
1999	19	13	3.08	0.84	0.97	0.968	1	NA	NA	
2000 2001	21 43	10 10	2.57 3.00	0 0.67	0 0.71	0.974	0.978	344 1	146 1	
2001	45 10	10	2.90	0.67	0.71	0.982	0.978	338	163	
2002	10	10	3.04	0	0	0.967	0.989	1212	567	
2004	17	10	3.18	0	0	0,97	0.97	1192	447	
2005	17	10	3.31	0	0	0.924	0.967	585	182	
2006	23	10	3.24	0	0	0.93	0.988	857	326	
2007	25	10	2.98	0	0	0.965	0.956	917	206	
2008	29	10	3.08	0	0	0.974	0.988	1593	260	
2009	28	10	3.50	0	0	0.934	0.894	1483	138	
2010 2011	43 27	10 10	3.42 3.40	0	0	0.982 0.939	0.975 0.831	991 888	249	
2011	35	10	3.40	0	1	0.939	0.851	1121	72 252	
2012	48	10	3.35	ů	1	0.949	0.966	938	211	
2014	70	10	3.32	0	1	0.913	0.961	660	260	
2015	61	10	3.37	0	1	0.97	0.982	1337	337	
	30	10		0	1	0.935	0.955	864	438	

Appendix 4: Input data for Atlantic salmon used to do the runreconstruction and estimates of returns and spawners by size group and age group for North America

Appendix 4.i. Input data for the fishery at West Greenland used in the run-reconstruction model.

Year of the fishery	Harvest in tonnes of salmon at West Greenland for assessment	Unreported harvest in tonnes at West Greenland		Prop. of harvest of NAC origin (min based on scales)	Prop. of harvest of NAC origin (max based on scales)	Prop. of NAC origin salmon that are 1SW	Prop. of NEAC origin salmon that are 1SW	Number of genetic samples that are NAC origin	Number of genetic samples that are NEAC origin
Year	WG.harv	WG.unharv	WG.meanwt		WG.propNAC_max		WG.prop1SWNEAC	WG.sampleNAC	WG.sampleNEAC
1971	2689				0.40	0.945			
1972	2113	C	3.44	0.34	0.37	0.945	0.964	C	(
1973	2341	C			0.59	0.945			
1974	1917				0.46	0.945		C	
1975	2030	C	3.12	0.40	0.48	0.945	0.964	C	
1976	1175				0.48	0.945			
1977	1420			0.38	0.57	0.945			
1978	984				0.57	0.945	0.964	C	
1979	1395	C	3.34	0.48	0.52	0.945	0.964	C	(
1980	1194	C	3.22	0.45	0.51	0.945	0.964	C	
1981	1264				0.61	0.945		C	
1982	1077			0.60	0.64	0.945		C	
1983	310	C		0.38	0.41	0.945		C	
1984	297			0.47	0.53	0.945			
1985	864			0.46	0.53	0.925			
1986	960				0.66	0.951	0.975		
1987	966				0.63	0.963	0.980		
1988	893				0.49	0.967		0	
1989	337			0.52	0.60	0.923		0	
1990	274				0.79	0.957	0.963	0	
1991	472				0.69	0.956	0.934		
1992	237			0.50	0.57	0.919	0.975		
1993	0				0.76	0.95			
1994	0				0.76	0.95			
1995	83				0.72	0.968			
1996	92				0.76	0.941	0.961	0	
1997	58			0.75	0.84	0.982			
1998	11				0.84	0.968	0.994		
1999	19				0.97	0.968	1.000		
2000	21					0.974			
2000	43				0.71	0.982			
2002	9.8				0.71	0.973			
2002	12.3				0	0.967			
2003	17.2					0.970			
2005	17.3			0		0.924			
2005	23.0					0.930			
2000	24.8					0.965			
2008	28.6					0.974			
2008	28.0					0.934			
2009	43.1					0.982			24
2010	27.4			0		0.939		888	
2011	34.6			0	0	0.935			
2012	47.7				0	0.932			
	47.7	10	3.30						
	55.0	10	3.32	0					
2013 2014 2015	4/// 552 55.9	10	3.35 3.32 3.37		0	0.949 0.913 0.970	0.961	660	1

Appendix 4.ii. Input data for sea fisheries on large salmon and small salmon from Newfoundland and Labrador used in the run reconstruction model. Labrador represents harvests from Labrador in aboriginal fisheries for food, social and ceremonial purposes and the resident food fishery beginning in 1998.

		Harvests of la	arge salmo	n (number)	Harvests of s	mall salmon (number)
			Salmon				
		Salmon	fishing	Labrador	Salmon	Salmon	Labrador
		fishing		subsistence	fishing	fishing areas	
					0		
		areas 1 to 7	14a	fisheries	areas 1 to 7	8 to 14a	fisheries
		Nlg_LBandN				Nsm_NF8to1	
Year		F1to7	o14a	NIg_LBFSC	NF1to7	4a	Nsm_LBFSC
	1970		0	0		0	
	1971		0	0			
	1972			0		111141	
	1973			0		176907	
	1974	196726	85714	0	192195	153278	
	1975	215025	72814	0			
-	1976	210858	95714	0	221766	118779	
-	1977	231393	63449	0	220093	57472	
-	1978	155546	37653	0	102403	38180	
-	1979	82174	29122	0	186558	62622	
	1980	211896	54307	0	290127	94291	
2	1981	211006	38663	0	288902	60668	
:	1982	129319	35055	0	222894	77017	
	1983	108430	28215	0	166033	55683	
	1984	87742	15135	0	123774	52813	
	1985	70970		0			
	1986	107561	22036	0		91912	
	1987	146242	19241	0		82401	
	1988						
	1989	85319		0		60884	
	1990	59334		0		46053	
	1991	39257	10259	0		42721	
	1992	32341		0		0	
	1993	17096		0			
	1994			0		0	
	1995			0		0	
	1995 1996		0	0		0	
	1997			0		0	
	1998						
	1998	0		2269			
		0		1084		0	
	2000			1352			
	2001	0		1721		0	
	2002			1389			
	2003			2175			
	2004						
	2005						
	2006						
	2007						
	2008						
	2009						
	2010						
	2011	0	0			0	
	2012		0			0	
2	2013	0	0	6479	0	0	718
2	2014	0	0	3994	0	0	895
2	2015	0	0	6146	0	0	892

Year of the	Reported harvest	Number of		Number of small
fishery	(kg)	salmon	large salmon	salmon
1970	0	0	0	0
1971	0	0	0	0
1972	0	0	0	0
1973	0	0	0	0
1974	0	0	0	0
1975	0	0	0	0
1976	3000	1331	333	998
1977	0	0	0	0
1978	0	0	0	0
1979	0	0	0	0
	0	-		
1980	-	0	0	0
1981	0	0	0	0
1982	0	0	0	0
1983	3000	1331	333	998
1984	3000	1331	333	998
1985	3000	1331	333	998
1986	2500	1109	277	832
1987	2000	887	222	665
1988	2000	887	222	665
1989	2000	887	222	665
1990	1900	843	211	632
1990	1200	532	133	399
	2300			
1992	2300	1020	255	765
1993		1287	322	965
1994	3400	1508	377	1131
1995	800	355	89	266
1996	1600	710	177	532
1997	1500	665	166	499
1998	2300	1020	255	765
1999	2322	1030	258	773
2000	2267	1006	251	754
2001	2155	956	239	717
2002	1952	866	217	650
2003	2892	1283	321	962
2004	2784	1235	309	926
2004	3287	1458	365	1094
2005	3555	1456	303	1183
2006	1947	864	394 216	648
2007	3540	1571	393	1178
2009	3460	1535	384	1151
2010	2780	1233	308	925
2011	3757	1667	417	1250
2012	1450	643	161	482
2013	5300	2351	588	1764
2014	3810	1690	423	1268
2015	3510	1557	500	1057

Appendix 4.iii. Input data for sea fisheries on large salmon and small salmon from St Pierre & Miquelon used in the run-reconstruction model.

	Large Salmon																	
	Commercial h	arvest (number	of fish)	FSC Fishery	Proportion La	abrador origin					Exploitation ra	ate	Proportion 2S	N	Returns to Labra	ador rivers	Angling catche)S
	SFA 1	SFA 2	SFA 14B	ALL	SFA 1		SFA 2		SFA 14B		All SFAs				number of fish		number of fish	
Year				Number of fish	Min	Max	Min	Max	Min	Max	Min	Max	: Min	Max	Min	Max	Retained	Released
1970	25127	64806	13673		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	0	562	0
1971	21599	55708	11753		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	0	486	0
1972	30204	77902	16436		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	0	424	0
1973	13866	93036	15863		0.6	0.8	0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	0	1009	0
1974	28601	71168	14752		0.6	for a second	,	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	0	803	0
1975	38555	77796	15189		0.6			0.8	0.6	0.8	0.7	0.9	0.70	0.90	0		327	0
1976	28158	70158	18664		0.6	2		0.8	0.6	0.8	0.7	0.9	0.70	0.90	0		830	0
1977	30824	48934	11715		0.6	f		0.8	0.6	0.8		0.9	0.70	0.90	0		1286	0
1978	21291	27073	3874		0.6			0.8	0.6	0.8	0.7	0.9	0.70	0.90	0		767	0
1979	28750	87067	9138		0.6			0.8	0.6	0.8	0.7	0.9	0.70	0.90	0		609	0
1980	36147	68581	7606		0.6	2		0.8	0.6	0.8	0.7	0.9	0.70	0.90	0	~	889	0
1981	24192	53085	5966		0.6		(0.8	0.6	0.8	0.7	0.9		0.90	0		520	0
1982	19403	33320	7489		0.6			0.8	0.6		0.7	0.9	0.70	0.90	0		621	0
1983	11726	25258	6218		0.6		0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0		428	0
1984	13252	16789	3954		0.6		0.6	0.8	0.6	0.8	0.7	0.9	0.70	0.90	0		510	0
1985	19152	34071	5342		0.6			0.8	0.6		0.7	0.9		0.90	0		294	0
1986	18257	49799	11114		0.6	2		0.8	0.6	0.8	0.7	0.9		0.90	0		467	0
1987	12621	32386	4591		0.6			0.8	0.6	0.8	0.7	0.9		0.90	0	×	633	0
1988	16261	26836	4646		0.6			0.8	0.6		0.7	0.9		0.90	0		710	0
1989	7313	17316	2858		0.6			0.8	0.6		0.7	0.9		0.90	0		461	0
1990	1369	7679	4417		0.6			0.8	0.6		0.7	0.9		0.90	0	*****	357	0
1991	9981	19608	2752		0.6			0.8	0.6		0.580	0.830		0.90	0		93	0
1992	3825	9651	3620		0.6			0.8	0.6	0.8	0.38	0.62		0.90		0	781	10
1993	3464	11056	857		0.6	3		0.8	0.6	0.8	0.29	0.50		0.90	0	0	378	91
1994	2150	8714	312		0.6		0.6	0.8	0.6	0.8	0.14	0.25		0.90	0	0	455	347
1995	1375 1393	5479 5550	418 263		0.6	for an and the second s		0.8 0.9521	0.6 0.6	0.8	0.13	0.23	0.70 0.70	0.90	0	×	408 334	508 489
1996		0000	203		0.6433	0.7247	0.0639	0.9521	0.0	0.8	0.17			0.90	0	0		
1997 1998	6943 0	0		2269	1			1	1	1	0.17	0.30	0.60	0.71	7374	19486	158 231	566 814
1998	0	0	0	2269				1	1	1	0.17	0.30	0.60	0.71	8827	23328	320	931
2000	0	0	0	1352					1		0.17	0.30		0.71	12052	31850	262	1446
2000	0	0	0	1352			1	1	1		0.17	0.30	0.60	0.71	12032		338	1440
2001	0	0	0	1721			1		1		0.17	0.30		0.71	9076		207	978
2002	0	0	0	2175			1		1		0.17	0.30		0.71	6676		222	1326
2003	0	0	0	3696			1	1	1	1	0.17	0.30		0.71	10964		259	1520
2004	0	0	0	2817		1	1	1	1	1	0.17	0.30		0.71	11159		200	1290
2006	0	0	0	3090		1	1	'	1	1	0.17	0.30		0.71	12414		227	1133
2000	0	0	0	2652		1	1	1	1	1	0.17	0.30		0.71	11887		235	1133
2007	0	0	0	3909	1	1	1	1	1	1	0.17	0.30		0.71	14700		200	1461
2009	0	0	0	3344	1	1	1	1	1	1	0.2	0.4		0.70	18643		216	1219
2000	0	0	0	3725	1	1	1	1	1	1	0.2	0.4		0.70	10764		197	1080
2010	0	0	0	4451	1	1	1	1	1	1	0.2	0.4		0.70	30198		0	2233
2012	0	0	0	4228	1	1	1	1	1	1	0.2	0.4		0.70	19062		0	1072
2012	0	0	0	6479	1	1	1	1	1	1	0.2	0.4	· · · · · · · · · · · · · · · · · · ·	0.70	36859]	0	2433
2014	0	0	0	3994	1	1	. 1		1		0.2	0.4	0.60	0.70	36055	87989	0	1385
2015	0	0	0	6146	1	1	1	1	1	1	0.2	0.4		0.70	49662	127898	0	1028

Appendix 4.iv. Input data for large salmon for Labrador used in the run-reconstruction.

Appendix 4.v. Input data for small salmon for Labrador used in the run-reconstruction.

Year Number of fist 1970 19109 38359 11212 1971 14303 28711 8392 1972 3130 6282 1836 1973 9848 37145 9328 1975 17589 47468 13152 1976 17796 40539 11267 1977 17095 12535 4026 1977 17095 12535 4026 1977 17095 12535 4026 1978 9712 28808 7194 1979 22501 72485 8493 1980 21596 86426 6658 1981 18478 53592 7379 1982 15964 30185 3292 1983 11474 11695 2421 1984 15400 24499 7460 1985 17779 45321 8296 1986 13714 64351 11389 1	SFA 1 Min 0.6	jin SFA 2 Max Mir 0.8 0.6	Max 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	SFA 14B Min 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	Max 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	Ate Max 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5	Returns to Labre number of fish Min 0	Max 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	Angling catche number of fish Retained 4013 3934 2947 7492 2501 3972 5726 4594 2691 4118 3800 5191 4104	
Year Number of fist 1970 19109 38359 11212 1971 14303 28711 8392 1972 3130 6282 1836 1973 9848 37145 9328 1974 34937 57560 19294 1975 17589 47468 13152 1976 17796 40539 11267 1977 17095 12535 4026 1978 9712 28808 7194 1979 22501 72485 8493 1980 21596 86426 6658 1981 18478 53592 7379 1982 15964 30185 3292 1983 11474 11695 2421 1984 15400 24499 7460 1985 17779 45321 8296 1986 13714 64351 11389 1987 19641 56381 7087	Min M 0.6 0.6	Max Mirr 0.8 0.6	Max 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	Min 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	Min 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5	Min 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0		Retained 4013 3934 2947 7492 2501 3972 5726 4594 2691 4118 3800 5191	
$\begin{array}{c c c c c c c c c c c c c c c c c c c $	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5	0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0		4013 3934 2947 7492 2501 3972 5726 5726 4594 2691 4118 3800 5191	Released () () () () () () () () () () () () ()
$\begin{array}{c c c c c c c c c c c c c c c c c c c $	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5	0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0		3934 2947 7492 2501 3972 5726 4594 2691 4118 3800 5191	
1972 3130 6282 1836 1973 9848 37145 9328 1974 34937 57560 19294 1975 17589 47468 13152 1976 17796 40539 11267 1977 17095 12535 4026 1978 9712 28808 7194 1979 22501 72485 8493 1980 21596 86426 6668 1981 18478 53592 7379 1982 15964 30185 3222 1983 11474 11695 2421 1984 15400 24499 7460 1985 17779 45321 8296 1986 13714 64351 11389 1987 19641 56381 7087 1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 <td>0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6</td> <td>0.8 0.6 0.8 0.6</td> <td>0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8</td> <td>0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6</td> <td>0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8</td> <td>0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3</td> <td>0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5</td> <td>0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0</td> <td></td> <td>2947 7492 2501 3972 5726 4594 2691 4118 3800 5191</td> <td></td>	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5	0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0		2947 7492 2501 3972 5726 4594 2691 4118 3800 5191	
$\begin{array}{ c c c c c c c c c c c c c c c c c c c$	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5	0 0 0 0 0 0 0 0 0 0 0 0		7492 2501 3972 5726 4594 2691 4118 3800 5191	
$\begin{array}{c c c c c c c c c c c c c c c c c c c $	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5	0 0 0 0 0 0 0 0 0		2501 3972 5726 4594 2691 4118 3800 5191	
$\begin{array}{c c c c c c c c c c c c c c c c c c c $	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5	0 0 0 0 0 0 0 0	0 0 0 0 0 0 0 0	3972 5726 4594 2691 4118 3800 5191	
$\begin{array}{c c c c c c c c c c c c c c c c c c c $	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5	0 0 0 0 0 0	0 0 0 0 0 0	5726 4594 2691 4118 3800 5191	
$\begin{array}{c c c c c c c c c c c c c c c c c c c $	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5 0.5 0.5	0 0 0 0 0	0 0 0 0 0	4594 2691 4118 3800 5191	0 0 0 0 0 0
1978 9712 28808 7194 1979 22501 72485 8493 1980 21596 86426 6658 1981 18478 53592 7379 1982 15964 30185 3292 1983 11474 11695 2421 1984 15400 24499 7460 1985 17779 45321 8296 1986 13714 64351 11389 1986 13744 56381 7087 1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5 0.5	0 0 0 0 0	0 0 0 0 0	2691 4118 3800 5191	C 0 0 0 0 0 0
$\begin{array}{c c c c c c c c c c c c c c c c c c c $	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5 0.5	0 0 0 0	0 0 0 0	4118 3800 5191	0 0 0 0
1980 21596 86426 6658 1981 18478 53592 7379 1982 15964 30185 3292 1983 11474 11695 2421 1984 15400 24499 7460 1985 17779 45321 8296 1986 13714 64351 11389 1987 19641 56381 7087 1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3 0.3 0.3	0.5 0.5 0.5 0.5	0 0 0	0 0 0	3800 5191	C C O
1981 18478 53592 7379 1982 15964 30185 3292 1983 11474 11695 2421 1984 15400 24499 7460 1985 17779 45321 8296 1985 17779 45321 8296 1986 13714 64351 11389 1987 19641 56381 7087 1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6 0.6	0.8 0.8 0.8 0.8	0.3 0.3 0.3 0.3	0.5 0.5 0.5	0 0	0	5191	C 0
1982 15964 30185 3292 1983 11474 11695 2421 1984 15400 24499 7460 1985 17779 45321 8296 1986 13714 64351 11389 1987 19641 56381 7087 1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753 9753 9793 1998 0 0 0 2739 2000 0 0 0 5328	0.6 0.6 0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8 0.8	0.6 0.6 0.6 0.6	0.8 0.8 0.8	0.3 0.3 0.3	0.5 0.5	0	0		0
1983 11474 11695 2421 1984 15400 24499 7460 1985 17779 45321 8296 1986 13714 64351 11389 1987 19641 56381 7087 1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753 9753 9799 1998 0 0 0 2739 2000 0 0 0 5323	0.6 0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6	0.8 0.8 0.8 0.8	0.6 0.6 0.6	0.8 0.8	0.3 0.3	0.5			4104	
1984 15400 24499 7460 1985 17779 45321 8296 1986 13714 64351 11389 1987 19641 56381 7087 1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753 9753 998 1998 0 0 0 2739 2000 0 0 0 5323	0.6 0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6 0.8 0.6 0.8 0.6	0.8 0.8 0.8	0.6 0.6	0.8	0.3		0			0
1985 17779 45321 8296 1986 13714 64351 11389 1987 19641 56381 7087 1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753	0.6 0.6 0.6 0.6	0.8 0.6 0.8 0.6 0.8 0.6	0.8 0.8	0.6			0 F		0	4372	0
1986 13714 64351 11389 1987 19641 56381 7087 1988 13233 34200 9053 1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753	0.6 0.6 0.6	0.8 0.6 0.8 0.6	0.8		0.8		0.5	0		2935	0
1987 19641 56381 7087 1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753	0.6 0.6	0.8 0.6	for the second s	0.6	0.0	0.3	0.5	0	0	3101	0
1988 13233 34200 9053 1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753	0.6		0.8		0.8	0.3	0.5	0		3464	0
1989 8736 20699 3592 1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753 9753 998 1998 0 0 2739 2000 0 0 5323		0.8 0.6		0.6	0.8	0.3	0.5	0	0	5366	0
1990 1410 20055 5303 1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753			0.8	0.6	0.8	0.3	0.5	0	0	5523	0
1991 9588 13336 1325 1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753		0.8 0.6	0.8	0.6	0.8	0.3	0.5	0		4684	0
1992 3893 12037 1144 1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753		0.8 0.6		0.6	0.8	0.3	0.5	0		3309	0
1993 3303 4535 802 1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753	0.6	0.8 0.6		0.6	0.8		0.39	0		2323	0
1994 3202 4561 217 1995 1676 5308 865 1996 1728 8025 332 1997 9753		0.8 0.6		0.6	0.8	0.13	0.25	0	f	2738	251
1995 1676 5308 865 1996 1728 8025 332 1997 9753		0.8 0.6	0.8	0.6	0.8	0.10	0.19	0		2508	1793
1996 1728 8025 332 1997 9753		0.8 0.6	0.8	0.6	0.8	0.07	0.13	0		2549	3681
1997 9753 1998 0 0 2988 1999 0 0 0 2739 2000 0 0 0 5323	elegence and a second	0.8 0.6	*******	0.6	0.8	0.04	0.07	0	ç	2493	3302
1998 0 0 2988 1999 0 0 0 2739 2000 0 0 0 5323		163 0.748		0.6	0.8	0.05	0.08	0	5	2565	3776
1999 0 0 2739 2000 0 0 0 5323	1	1 1	furner and the second	1	1	0.05	0.08	0		2365	2187
2000 0 0 5323		1		1	1	0.05	0.08	97408	7	2131	3758
	1	11	f	1	1	0.05	0.08	94894	199901	2076	4407
2001 0 0 0 4789	1	1 1		1	1	0.05	0.08	117063	246602	2561	7095
		1 1		1	1	0.05	0.08	93660	197301	2049	4640
	1	1 1	******	1	1	0.05	0.08	62321	142951	2071	5052
2003 0 0 0 6477	1	1 1	f	1	1	0.05	0.08	48256	122813	2112	4924
2004 0 0 0 8385	1	1 1		1	1	0.05	0.08	69808	fananananananananananananan	1808	5968
2005 0 0 0 10436	1	1 1		1	1	0.05	0.08	160038		2007	7120
	1	1 1		1	1	0.05	0.08	132205		1656	5815
2007 0 0 9208	1	1 1		1	1	0.05	0.08	131895		1762	4641
2008 0 0 0 9834	1	1 1		1	1	0.05	0.08	142851	264694	1936	5917
2009 0 0 7988	1	1 1		1	1	0.07	0.14	55307	149372	1355	3396
2010 0 0 9867	1	1 1			1	0.07	0.14	78560	165165	1477	4704
	1	1 1		1	1	0.07	0.14	137465		1628	5340
2012 0 0 0 9977	****	1 1	i	1	1	0.07	0.14	105443	5	1376	3302
2013 0 0 7185	1	1 1		1	1	0.07	0.14	83556	227719	1389	4167
2014 0 0 8958 2015 0 0 0 8923	****	1 1	1	1	1	0.07 0.07	0.14	175938 174788	5	1370 1139	3797 3970

Appendix 4.vi. Input data for returns of small salmon and large salmon for Salmon Fishing Areas	as 3 to 8 in Newfoundland used in the run-reconstruction.
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		Salmon Fishin	g Area 3			Salmon Fis	hing Area	4	5	Salmon Fis	shing Are	ea 5		Salmon Fi	shing Are	a 6		Salmon Fi	shing Are	a 7	S	almon F	ishing Ar	ea 8
	Smal	salmon	ř	salmon		salmon	Large s			salmon		e salmon		salmon		salmon		salmon	- ×	e salmon	Small s		-	e salmon
	Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Return	,
Year	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
Bugs labe	SFA3Sm_L[]	SFA3Sm_H[]	SFA3Lg_L	[SFA3Lg_H	SFA4Sm_	SFA4Sm_	SFA4Lg_L	SFA4Lg_H	SFA5Sm_	SFA5Sm_	SFA5Lg	L SFA5Lg	FSFA6Sm	SFA6Sm	SFA6Lg	L SFA6Lg_I	SFA7Sm	SFA7Sm	_SFA7Lg	L SFA7Lg_	SFA8Sm_	SFA8Sm	SFA8Lg	_L SFA8Lg_
1970	261	3 5227	155	5 73	7 16163	32327	957	4559	7420	14840) 4	139 209	3 280	0 56	0 1	7 79	6	7 13	3	4 1	9 62	12	23	4 1
1971	247	3 4947	7 146	5 69	8 12610	25220	746	3557	5600	11200) 3	331 157	9 183	3 36	7 1	1 52	2 13	3 26	7	8 3	8 83	10	57	5 2
1972								3238				374 178			3 2	23 112				12 5		18		6 2
1973								6308				17 198				19 235				26 12		62		19 8
1974								4131	5457			529 125								51 10		34		20 3
1975								5587				035 186			-	14 88				19 3		58		41 8
1976								5365				62 152				9 198				12 2		53		32 6
1977								5290				154 290				26 253				25 4			10	26 5
1978								3461				564 112				58 117				20 3			03	9 1
1979							1533	3067				254 50				17 93				24 4		60		19 3
1980								2819				405 80				18 96				15 3			00	18 3
1981								11996				546 309				8 530				55 10			13	34 6
1982							802	1604				196 39				3 40				12 2			57	7 1
1983 1984								2740				361 72			· · · · · · · · · · · · · · · · · · ·	15 91 17 168				21 4 5 5			73	6 1 4 4
								5337				150 145 192 181		-		17 168 24 224				5 5 6 5		7	94	4 4 6 5
1985							671 840	6352 4977				192 181 366 217				24 224 10 238				9 5		105		12 74
1980								3078				170 93		0 108		4 77				3 1			10	4 2
1987								5367				377 226		333		39 234				12 7		6		7 4
1989							461	2365				72 88		203		30 154				8 4		82		12 6
1900							776	3033				331 129				6 179				7 2		59		12 0
1991				-				2788				01 116				27 106				5 2			78	2
1992								13867	14189			516 508				51 603				1 10			0	0
1993							1605	6242				586 228				01 352				16 6			3	15 5
1994								7343				73 243				32 135				4 1			13	7 2
1995								8211				63 290				2 10				14 6			16	11 4
1996								11497								7 14				1 3			00	16 4
1997								8125			5	281 380 574 218				25 79				6 2			0	6 2
1998	3 1961	7 27571			2 52347	73573	5219	18658	11863	16673		183 422	8 538	8 75	6 5	54 192	2 24	9 35	0 3	25 8	9 161	22	27	16 5
1999								18651	10474	15245		064 314				37 122				6 2			20	14 4
2000	1931	3 26033	3 1460	5 372	8 37551	50618	2850	7248	12414	16734	1 9	042 239	6 112	8 152	0 8	36 218	3 15	9 21	4	12 3	1 106	14	13	8 2
2001	1175	4 15383	3 90	7 210	4 39901	52218	3080	7143	10007	13095	5 7	73 179	1 290	5 38	7 2	23 53	3 5	3 6	9	4	9 20	1	26	2
2002	2 1050	0 15736	5 684	1 200	6 34310	51418	2234	6556	3870	5799	2	252 73	9 24	1 36	1 1	6 46	5 (0	0	0	0 72	10	08	5 14
2003	3 2161	5 26166	5 1092	2 348	5 74615	90328	3768	12032	6583	7970) 3	332 106	2 458	8 55	5 2	23 74	1 10	4 12	6	5 1	7 52	(53	3
2004	1 799	2 12452	396	5 1680	6 49598	3 77280	2455	10464	8385	13065	5 4	15 176	9 180	0 28	1	9 38	3	0	0	0	0 41	(54	2
2005	642	1 18899	487	267	8 36753	108180	2790	15329		15627	4	403 221	4 114	4 33	6	9 48	3	0	0	0	0 26	1	76	2 1
2006	6 1075	7 17194	125	323	9 42745	68322	4971	12872	8571	13700) 9	997 258	1 69	9 11	0	8 21	L I	0	0	0	0 172	27	75	20 5
2007		2 21117	1182	2 382	8 36934			13567	8734	17696	5 9	90 320	8 78	8 15	7	9 28		9 26	2	15 4	7 17	3	35	2
2008		1 23285	5 1062	2 339				15508	11459	19195	5 8	376 280	0 330	0 55	2 2	25 81			1	6 2	1 196		29	15 4
2009								22760				527 405				29 185			0		0 135		52	8 5
2010							6094	17393						7 124		6 218				6 4		13		8 24
2011							5033	24165								1 339				8 4		46		23 10
2012								13600						7 117		8 155				8 2	1 408	58	0	29 7
2013								23377	17957							8 376				6 12		18		7 5
2014							5170	13266								5 168				25 6		15		8 2
2015	5 2084	4 35159	9 1717	514	4 83603	141016	6886	20631	17154	28934	14	13 423	3 1158	3 195	2 9	5 286	5 27	9 47	0 2	3 6	9 357	60	12	29 8

	S	almon Fis	ning Area	9	S	almon Fish	ning Area 10		S	almon Fis	hing Area 1	1	S	almon Fis	hing Area	12	Sa	almon Fis	hing Area	13		Salmon Fish	ing Area 14A	
	Small s	salmon	Large s	salmon	Small	salmon	Large salm	non	Small	salmon	Large s	almon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon
	Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns	
Year	Min	Max	Min	Max	Min	Max	Min Max	x	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
lugs labe	SFA9Sm_	SFA9Sm_	SFA9Lg_L	SFA9Lg_H	SFA10Sm	SFA10Sm	SFA10Lg_SFA	A10Lg_	SFA11Sm	SFA11Sm	SFA11Lg_	SFA11Lg_	SFA12Sm	SFA12Sm	SFA12Lg	SFA12Lg	SFA13Sm	SFA13Sm	SFA13Lg_	SFA13Lg_	SFA14ASm_	SFA14ASm_H	SFA14ALg_L	SFA14ALg_H
1970		12620	373	1780		4007	119	565	16760	33520		4727	2497					38282		5060				
1971	5400	10800	320	1523	3093	6187		872	13533	27067	801	3817	1513	3027	90	427	26011	40151	2678	4750	12523	25047	308	2173
1972		7593	225	1071	1890	3780		533	16350	32700		4611	3093					37589		5169				
1973		14400	426	2031	5950	11900		1678	16187	32373		4565	2153					40227		5200			433	
1974		9960	574	1149		8080		932	14920	29840		3441	2193					28824		4257			902	
1975		12480	880	1760		2847	201	401	15003	30007		4232						54424		7424			507	
1976		10820	651	1303		4867		586	13880	27760		3343					29382	46902					1437	
1977	3600	7200	340	680		7313		691	13653	27307		2581	1860			352		25240		3598			666	
1978		8687	257	514		10633		629	13320	26640		1576						27681					266	
1979		11360	326	651		5660		324	11433	22867		1311	2443			280		31829		1862			233	
1980		15860	356	712		10160		456	16897	33793		1518						38871					694	
1981	6207	12413	825	1650				1167	23540	47080		6258	3533		470		31359	45989					1090	
1982		12167	147	293		8373		202	24460	48920		1180					31628	46698			15877		3094	6189
1983		15353	352	705				349	15897	31793		1460						31701					1704	
1984		17023	125	1221	5141	10955		785	24767	52774		3784	6782					37852		3710				
1985		13198	96	912		10000		691	21213	43914		3034	3996			572		25505		2508				
1986	8411	17555	208	1231	5619	11727	139	822	20300	42368		2970						36916		3649			455	
1987	3416	6865	88	489		3397	44	242	15087	30317		2162			85			32325		3022			471	
1988		10668	124	748		8873		622	18985	39106		2741	5330		128			43480					566	
1989		10895	160	821	3655	7440		561	12047	24524		1849	2279		68			16156		1565			390	
1990		12834	256	1000		5743		447	17470	30578		2382	3363		117			31183					693	
1991	2404	3949	82	319		1622		131	7956	13068		1056						20945		2433			523	
1992		10088	184	1809		3582		642	16615	33231		5958		9342				44119		8928			1271	12525
1993		21948	401	1560		10736		763	24574	47301		3363	5936					46889		4746			1095	
1994	3007	6607	177	751	2544	5588		635	7649	16803		1910	2761					37166					783	
1995		12821	318	1390		10532		1142	10757	25916		2809	2294					49781		5096			1244	
1996 1997		13450	443 379	1317				1805	18938	42350		4146 5619	5025 4556					67672 46494					2415 2630	
1997	3636 4694	6627 6597	468	1227		9242		1712 2788	16648 8467		844	3018								7080				
						10992				11900	844							27955					2392	
1999		5844	369	1205				1535	9643	14036	88/	2894	1139					40858		7739			2480	
2000 2001	7850 2043	10582 2674	596 158	1515		10297 3826		1475 523	17260	23266 12296	1310 725	3332 1682						67784					2795	
2001		26/4 2873		366				523 709		12296								45761					1604	
2002	2229	28/3	125	366		5565 4565		609	9011			1722 2291	2321					46011		8434			2061	5075
2003	1926	2099	95	339 406		5760		780	14208 13762	17201 21443		2291	3131					66549		9118			2001	
2004	1926	5734	95	813		8180		1150	6260			2903	2686					88340		12966			1488	
2005	4355	5734 6960	506	1311		8180		1609	11033			3322	2080					74546		12966			3266	
2006	2377	4817	270	873		7086		1285	5650	17634		2076						59140		15058			2264	
2007		6606	301	963		8016		1285		18654		2070	2610					75122		9668			1972	
2008		6443	203	1316		9608		1963	7536	14097		2721	1746					55458					1972	
2009		8227	203 504	1316		10187	625	1783		14097		2880	2999					67116		110806			2422	
2010	5271	8983	438	2105			750	3600	6897	11755		2754	2999					68766						
2011		9539	436	1259		11960		1579	6727	9554		1261	2469					71123		9390			2005	
2012		7015	266	2083	8060	11960		3527	7372	10863		3226	2024					49737					2465	
2013		7015	403	2083	5934	8933	450	1193	4408	6637		3226	1754					33219		5928			1392	
2014		8089	403 395	1183		8235		1205	6057	10217		1495	2167					84630		14188			2674	
2015	4795	0069	395	1103	4003	0235	402	1205	0057	10217	499	1495	2107	3000	1/0	035	50174	04030	0027	14100	32403	04/50	20/4	8011

Appendix 4.vi. (Continued). Input data for returns of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

Appendix 4.vii. Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

		Salmon Fishing	Area 3			Salmon Fi	shing Area	4		Salmon Fi	shing Area	5		Salmon F	ishing Area	16		Salmon F	ishing Area	a 7		Salmon F	ishing Are	a 8
	Small	salmon	Large	salmon	Smal	l salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	e salmon	Smal	l salmon	Larg	e salmon	Small	salmon	Larg	e salmon
	Spawners		Spawners		Spawners	S	Spawners		Spawners		Spawners	5	Spawners	3	Spawner	s	Spawner	S	Spawner	rs	Spawners	3	Spawne	rs
Year	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
	SFA3SSm_L[SFA3SSm_H[]	SFA3SLg	SFA3SLg	SFA4SS	m SFA4SSr	n SFA4SLg	_SFA4SLg	SFA5SSm	SFA5SSr	n SFA5SLg	SFA5SL	SFA6SSr	m SFA6SS	m SFA6SL	g_SFA6SL	g_SFA7SS	m SFA7SS	m SFA7SL	g_SFA7SL	g_SFA8SSI	m SFA8SS	m SFA8SL	.g_SFA8SLo
1970	1829	4443	154	736	11314	27478	910	4512	5194	12614	404	2058	196	476	14	76	47	113	3	18	43	105	0	13
1971	1731	4205	135	687	8827	21437	688	3499	3920	9520	293	1541	128	312	10	51	93	227	8	38	58	142	0	15
1972	1162	2822	98	468	8036	19516	655	3214	4422	10738	354	1762	278	674	23	112	142	346	12	57	65	159	6	26
1973	2772	6732	232	1115	15657	38023	1275	6259	4928	11968	405	1974	583	1417	49	235	306	742	26	123	219	533	15	84
1974	1958	4754	318	641	12537	30447	1983	4049	3820	9276	608	1237	707	1717	115	232	310	754	49	100	119	289	20	39
1975	2583	6273	520	1041	13867	33677	2628	5421	4639	11265	912	1846	219	533	43	87	93	227	19	38	203	493	41	82
1976	2210	5366	379	759	15594	37870	2495	5177	4429	10755	697	1459	576	1400	-97	196	70	170	12	24	187	453	32	64
1977	3570	8670	478	960	19591	47577	1559	4204	10771	26157	1410	2864	936	2272	107	234	182	442	24	48	189	459	26	51
1978	1769	4295	149	298	20473	49719	1229	2959	6669	16195	536	1100	691	1677	51	110	231	561	19	38	103	249	9	17
1979	4760	11560	390	779	18727	45481	1206	2740	3106	7542	234	489	569	1383	45	91	292	708	24	48	233	567	19	38
1980	4067	9877	224	485	21966	53346	903	2312	6305	15311	376	780	747	1813	34	82	238	578	14	30	280	680	18	36
1981	5502	13362	1042	2087	31584	76704	5637	11635	8139	19765	1511	3056	1412	3428	239	507	287	697	53	107	180	436	34	68
1982	6146	14926	124	336	23270	56514	544	1346	5677	13787	143	338	672	1632	6	29	362	878	2	15	198	482	0	5
1983	3773	9163	245	493	20893	50739	1073	2443	5500	13356	191	551	691	1677	35	81	324	788	0	9	96	232	1	8
1984	2531	6525	55	540	25033	64536	533	5322	6835	17620	149	1456	789	2034	12	163	243	626	1	48	200	515	4	43
1985	3462	8569	72	683	32218	79741	671	6352	9208	22791	192	1816	1134	2806	24	224	296	733	6	58	272	674	6	54
1986	2054	5126	70	413	24722	61700	840	4977	10782	26910	366	2170	1184	2955	40	238	271	677	9	55	367	916	12	74
1987	1655	3895	57	318	16032	37722	556	3078	4892	11511	170	939	403	948	14	77	82	194	3	16	126	297	4	24
1988	4868	11888	159	956	27317	66712	892	5367	11549	28204	377	2269	1189	2904	39	234	355	867	12	70	219	535	7	43
1989	2266	5376	90	461	11623	27581	461	2365	4350	10323	172	885	755	1792	30	154	203	481	8	41	304	721	12	62
1990	5032	10098	236	920	16583	33273	776	3033	7071	14188	331	1293	978	1963	46	179	144	288	7	26	252	505	12	46
1991	4334	7965	193	750	16113	29607	718	2788	6745	12395	301	1167	-613	1126	27	106	119	218	5	21	36	67	2	6
1992	9844	21262	415	4094	33228	71898	1407	13866	12175	26363	516	5088	1450	3132	61	603	252	545	11	105	0	0	0	0
1993	10054	20961	400	1599	39162	81344	1590	6226	14370	29779	576	2270	2243	4623	90	351	404	831	16	63	369	760	15	58
1994	9146	24802	749	3247	20576	55760	1644	7259	6855	18510	560	2420	381	1026	30	133	46	122	4	16	79	212	6	27
1995	7409	21791	580	2636	22872	67179	1801	8135	8122	23776	642	2880	287	831	23	100	173	501	14	60	135	397	11	48
1996	15729	39860	1412	4247	42346	107268	3757	11383	14095	35586	1263	3787	522	1317	46	139	124	311	11	33	180	457	16	48
1997	9422	19095	1209	3954	19309	39107	2467	8083	5228	10547	668	2177	190	384	24	79	49	99	6	20	48	98	6	20
1998	16390	24345	1933	6969	43559	64785	5160	18599	9943	14/53	947	4201	455	673	53	191	212	313	25	88	135	201	16	57
1999	11804	18173	1279	4189	52390	80698	5650	18583	8832	13603	2.11	3126	343	528	37	121	58	90	6 12	21	119	188	14	45
2000	17003	23723	1449	3711	32879	45946	2803	7201	10897 8344	15217	923	2377	993	1386	84	217	140	195	4	31	88	125	8	20
2001 2002	9861 8620	13489 13856	892	2089	33365 28099	45682	3023	7086	3194	11433 5124	250	1786 737	250	342 319	23	53	42	59	4	9	17	23	5	4
2002	19386	23938	671 1085	3478	67026	45208 82739	3738	6498 12001	5926	7312	331	1060	412	508	23	43	94	116	5	17	47	58	3	8
2003	6942	11402	390	1680	43104	70785		10438	7307	11987	412	1766	158	259	23	38	94	0	0	0	35	58	2	9
2004	5056	11402	390 473	2664	28896	100323	2430	10438	4200	14518	394	2205	92	314	8	38	0	0	0	0	18	58 69	2	11
2005	9402	17534	1228	3216	37156	62732	4925	15255	7495	14518	969	2205	61	102	8	20	0	0	0	0	18	244	20	52
2006	9402	19842	1228	3216	3/150	70143	4925	12825	7641	12623	969	3196	68	102	8	20	112	245	12	45	141	33	20	6
2007	11799	21183	1045	3379	53591	06142	4745	15301	9669	17405	867	2791	274	497	22	78	69	125	4	18	159	292	15	48
2008	11799	22795	779	5080	49881	101728	3491	22732	8828	18065	622	4049	412	834	22	185	09	0	0	0	111	292	7	51
2009	11205	22795	1595	4581	69075	88772	6006	17304	20114	25822	1754	5028	874	1120	76	217	183	235	16	46	93	120	8	24
2010	13193	24264	1393	6261	50806	93428	4789	23920	12075	23822	11754	5734	716	1314	70	339	83	153	8	39	220	412	22	108
2011	21149	31048	1639	4394	64959	95436	5046	13528	14554	21377	1140	3039	738	1086	57	154	100	133	8	21	361	533	25	73
2012	7822	12219	495	4039	44838	70136	2889	23284	15027	23531	976	7833	730	1128	44	373	241	379	16	127	102	162	7	55
2013	11442	18452	1074	2775	54396	87757	5068	13164	14519	23389	1345	3497	693	1115	62	165	255	416	24	63	83	134	8	21
2014	18245	32560	1690	5117	73398	130811	6786	20531	15040	26820	1392	4213	1019	1814	94	284	238	429	23	68	301	546	29	88

Appendix 4.vii. (Continued). Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

		Salmon Fi	ishing Area	9		Salmon Fi	ishing Area	10		Salmon Fis	hing Area	11		Salmon Fis	shing Area	12		Salmon Fis	shing Area 1	13		Salmon Fis	hing Area 14A	· · · · · · · · · · · · · · · · · · ·
	Small	salmon	Large	e salmon	Small	salmon	Large	e salmon	Small	salmon	Large	e salmon	Smal	l salmon	Large	salmon	Smal	l salmon	Large	salmon	Sma	II salmon	Larg	e salmon
	Spawners		Spawners		Spawners		Spawners		Spawners		Spawners		Spawners		Spawners		Spawners		Spawners		Spawners		Spawners	
Year	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
1[]	-			J_SFA9SL			SI SFA10SI						*			-				-				g_SFA14ASLg_
1970	4417	10727	361	1768	1402	3406	112	558	11732	28492	918	4653	1748	4244	69	625	16203	28543	1608	3417	10372	25188	134	2340
1971	3780	9180	301	1504	2165	5259	166	855	9473	23007	736	3752	1059	2573	74	411	16489	30629	1633	3705	8766	21290	0	1850
1972	2658	6454	217	1063	1323	3213	108	529	11445	27795	882	4525	2165	5259	163	852	15125	29188	2004	4066	5640	13696	83	1283
1973	5040	12240	406	2011	4165	10115	310	1636	11331	27517	923	4530	1507	3661	102	582	17019	29959	1911	3808	12325	29931	91	2713
1974	3486	8466	565	1140	2828	6868	452	918	10444	25364	1682	3403	1535	3729	240	493	12085	21635	1997	3341	7280	17680	789	1692
1975	4368	10608	874	1754	996	2420	192	392	10502	25506	2076	4192	1190	2890	220	459	21668	42421	3611	6538	11242	27302	417	925
1976	3787	9197	639	1291	1703	4137	283	576	9716	23596	1629	3301	693	1683	114	233	18999	36519	2752	4862	17222	41826	1337	2774
1977	2520	6120	331	671	2560	6216	341	686	9557	23211	1272	2563	1302	3162	128	304	10898	18528	1828	2549	13316	32340	194	859
1978	3040	7384	240	497	3722	9038	273	587	9324	22644	770	1558	854	2074	52	124	12518	22392	3861	4434	7562	18366	194	460
1979	3976	9656	311	636	1981	4811	154	316	8003	19437	648	1304	1710	4154	130	270	14363	25820	1070	1749	15349	37275	174	408
1980	5551	13481	295	651	3556	8636	201	429	11828	28724	715	1474	1913	4647	94	217	18625	30958	4243	4920	8734	21210	514	1208
1981	4345	10551	773	1598	3073	7463	555	1138	16478	40018	3088	6217	2473	6007	453	922	22059	36689	4485	6789	13725	33331	953	2043
1982	4258	10342	114	260	2931	7117	91	192	17122	41582	537	1127	3628	8812	110	235	22062	37132	2847	3236	11114	26990	2987	6082
1983	5374	13050	281	634	2660	6460	95	270	11128	27024	703	1433	1556	3780	94	196	14491	25364	3855	4490	8867	21533	1635	3338
1984	5725	14759	120	1216	3684	9498	79	783	17748	45755	374	3769	4860	12529	38	968	18413	30081	1987	3401	12155	31336	179	2504
1985	4625	11448	96	912	3505	8674	73	691	15390	38091	320	3034	2899	7176	57	569	10726	20203	1349	2482	9583	23719	197	1887
1986	6113	15257	208	1231	4084	10192	139	822	14754	36822	501	2970	2495	6228	81	499	15535	29570	2013	3583	13381	33396	445	2683
1987	2549	5998	88	489	1261	2968	44	242	11258	26488	391	2162	2443	5749	82	466	13611	26307	1512	2988	13583	31960	467	2604
1988	3806	9295	124	748	3166	7731	103	622	13952	34073	456	2741	3917	9566	126	767	17945	35263	1909	3877	17329	42319	549	3388
1989	4037	9580	160	821	2757	6542	109	561	9087	21564	360	1849	1719	4080	67	349	6980	12982	836	1552	9833	23334	385	1996
1990	5466	10968	256	1000	2446	4908	115	447	13024	26132	610	2382	2507	5031	114	456	14866	24531	1744	3051	14793	29682	679	2692
1991	1844	3389	82	319	758	1392	34	131	6103	11215	272	1056	2121	3898	93	365	11037	15757	1689	2413	11742	21576	512	2020
1992	4334	9378	183	1809	1496	3287	65	642	14239	30854	605	5958	3985	8657	162	1667	20506	38635	2992	8833	30096	65023	1234	12488
1993	9956	20502	400	1559	4809	9967	194	761	21423	44150	861	3359	5176	10666	207	810	22341	41708	2544	4673	27010	55787	1058	4221
1994	2124	5723	172	746	1804	4848	144	630	5295	14449	430	1891	1949	5253	154	681	15381	30444	3207	5611	9385	25327	742	3286
1995	3887	11386	304	1376	3218	9378	253	1133	7770	22930	625	2792	1689	4922	130	592	20570	43329	1607	4860	15218	44587	1187	5385
1996	4868	12304	431	1304	6687	16880	592	1789	15226	38638	1362	4113	4082	10295	358	1088	29056	60152	3199	6852	26584	67085	2357	7115
1997	2927	5918	372	1221	4086	8257	519	1702	13304	26995	1718	5602	3655	7401	464	1527	25508	40599	3985	8266	20359	41117	2578	8467
1998	3937	5840	458	1663	6606	9777	771	2779	7024	10457	836	3009	1968	2925	225	831	18279	24417	3031	6918	19992	29721	2347	8507
1999	3401	5230	359	1195	4313	6642	455	1520	8086	12478	881	2889	958	1477	102	339	28647	37098	3760	7621	22659	34941	2402	8013
2000	6913	9645	581	1501	6664	9322	534	1429	14895	20901	1288	3310	2291	3208	195	504	48055	61508	5250	9779	32314	45127	2731	7044
2001	1709	2339	151	359	2436	3338	215	513	7804	10704	714	1671	1818	2497	162	386	31037	39405	3536	6297	17331	23744	1559	3674
2002	1562	2518	118	360	3049	4901	231	699	7347	11840	581	1716	1896	3053	147	439	28083	40025	3313	6330	21764	35007	1668	5013
2003	1985	2454	109	355	3368	4162	185	603	12701	15693	703	2276	5282	6528	288	943	45027	52657	4206	8218	36597	45189	1988	6506
2004	1674	2749	91	402	3210	5273	177	774	11863	19544	660	2882	2704	4452	149	655	43889	60513	4074	8883	26116	42892	1429	6282
2005	1478	5264	130	794	2171	7572	194	1142	4827	16992	456	2591	2062	7282	191	1107	33349	81031	4320	12691	13676	47376	1246	7163
2006	3791	6397	498	1302	4627	7824	602	1590	9554	16155	1271	3310	2986	5056	392	1032	46296	67532	8247	14807	24532	41334	3210	8400
2007	2063	4502	263	867	3047	6636	387	1275	4907	10706	636	2071	2442	5323	314	1027	29402	54734	4511	10780	17446	37934	2222	7293
2008	3285	5948	293	955	3971	7202	351	1154	9314	16832	841	2711	2178	3940	193	631	43277	66465	3580	9346	21887	39305	1915	6246
2009	2835	5834	198	1311	4193	8665	298	1957	6203	12763	442	2877	1450	2970	100	664	31106	50196	3526	10610	17820	36229	1200	8032
2010	5703	7334	496	1432	7062	9081	616	1774-	6859	8842	604	1742	2606	3347	226	651	49703	58889	5478	10747	27468	35298	2358	6848
2011	4364	8077	433	2099	7477	13826	716	3566	5696	10554	564	2744	2074	3827	203	990	33849	62267	3160	15915	20249	37231	1953	9575
2012	5898	8720	471	1256	7488	11027	581	1566	5993	8819	468	1255	2348	3450	184	490	44778	65820	3395	9251	31467	46268	2451	6571
2013	3973	6228	254	2071	6681	10498	424	3502	6130	9621	398	3213	1701	2668	104	885	28314	44299	3301	13638	11746	18341	734	6050
2014	4174	6776	400	1031	4862	7861	431	1158	3602	5830	340	881	1445	2331	134	349	18039	29192	2707	5849	14639	23622	1361	3541
2015	4169	7462	385	1173	4265	7618	389	1192	5250	9410	489	1485	1891	3379	167	524	43908	78364	5419	13980	28414	50707	2619	7956

Appendix 4.viii. Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

		Salmon Fish	ing Area 3			Salmon Fis	shing Area 4			Salmon Fi	shing Area 5			Salmon Fi	shing Area 6			Salmon Fis	ning Area /	·	5	almon Fis	hing Area 8	8
		2SW	2	SW	25	SW	2	SW		2SW	25	SW	2	2SW	25	SW	2	SW	25	SW	2SV	V	25	SW
	Returns		Spawners		Returns		Spawners		Returns		Spawners		Returns		Spawners		Returns		Spawners		Returns		Spawners	
Year	Mi	in Max	Min	n Max	Min	Max	Min	Max	M	in Max		Max	Mir	n Max	k Min	Max	Mir	n Max		Max	Min	Max	Min	n I
gs labels	SFA3R2_L[]	SFA3R2_H[]	SFA3S2_L[] SFA3S2_H[]	SFA4R2_L[]	SFA4R2_H[]	SFA4S2_L[]	SFA4S2_H[]	SFA5R2_L	[] SFA5R2_H[]	SFA5S2_L[]	SFA5S2_H[]	SFA6R2_	L SFA6R2_H	[SFA6S2_L	SFA6S2_H	SFA7R2_	L SFA7R2_H	SFA7S2_L	SFA7S2_H	SFA8R2_L S	SFA8R2_H	SFA8S2_L	L SFA8S
1970	1	5 147	15	5 147	96	912	91	902	2 4	41 419	40	412		2 1	6 1	15	i I	0 4	0	4	0	3	0	3
1971	1	5 140		4 137	75	711	69	700		33 310						10		1 8	1	8	0	5		J
1972		0 94								37 35				2 2		22		1 11		11		5		1
1973		23 223					127	1252	2 4	42 39	7 40	395		5 4		47		3 25		25		18	1	1
1974		32 129								53 252								5 20		20		8	2	-
1975	5	52 208				1117	263	1084		37-	1 91	369		4 1	8 4	17		2 8	2	8	4	16		
1976		38 152								76 30				4		39		1 5	-	5	-	13		
1977		8 193								45 582				3 5	1 11	47		2 10				10		
1978		5 60								56 220				5 2	3 5	22		2 8	-	0	-	3		
1979		9 156								25 102				5 1		18		2 10	2	10		8	2	
1980	2									40 162				5 1		16		2 6	-	6		7	2	-
1981	10									55 618			2	10	7 24	101		5 22				14		
1982	2									20 78					9 1	6		1 5		2	-	3		
1983		99								36 14				5 1	8 4	16		2 9				3		
1984		6 108								15 29				2 3		33		1 10		10		9		
1985		7 137								19 363		363		2 4:				1 12		12		11	1	-
1986		7 83								37 434		434		4 4		48		1 11		11		15		•
1987		6 64								17 18				1 1		15		0 3		-		5	0)
1988		6 191								38 454				4 4		47		1 14		14		9	1	1
1989		9 92								17 17				3 3		31		1 8	-	8	-	12		1
1990		24 184								33 259		259		5 3				1 5		5		9	-	1
1991		9 150								30 23.		283		3 2				1 4	-	4		1	0	
1992		12 819								52 101		1018		6 12		121		1 21		21		0	0)
1993		12 323								59 450				9 7		70		2 13		1.0		12		1
1994 1995	4									34 34 40 40	34	339 403		2 1		19		0 2		2		4		1
1995		57 573 36 598								400 77 53	3 76			3 2		14				5	-	7	1	4
										40 30	5 40							1 5 0 3			-	3	0	۱ 0
1997 1998	11									40 300 71 592				1 1 3 2		27		1 12		12		3	-	1
1998	7									8 44				2 1		17		0 3				6	1	•
2000		38 522								27 22	5 55			5 3		30		1 4	0	4	•	3	0	1
2000	3									33 16					5 1	50		0 1	0		0	0		
2001										10				1				0 0	0	0	0	1	0	
2002	4									14 99					+ 1 7 1	4						1	0	
2003		7 157								14 97					4 0	4		0 2	0		0	1	0	•
2004	2							1417		17 20				0 . D .		4		0 0		0		1	0	
2005		54 301								13 24					2 0					0		5	1	1
2008	5									43 298					3 0			1 4		4	-	1	0	0
2007		6 316						1432		38 26					7 1			•			0	4		1
2000		34 473						0110		27 37						17					-		0	
2009								1609		76 46				3 2		20		1 4		4		2	0	
2010	5									2 535				3 3				0 4	-			10		
2011										9 283				3 14		14		0 2	0			7		1
2012	2									3 73'				2 3		35		1 12		12		5		<u>.</u>
2013							218			is 73				3 10		15		1 6		6		2		-
2014										5 520 51 394				4 2		26		1 6				8		
2010		- 4/0	13	4/0	290	1919	292	1909			00	392		- Z	4	20	·	0		0	I	0		

		Salmon Fis	shing Ale	a 9		58	umon Fish	ing Area 10			Salmon Fi	shing Area 11				ing Area 12			Salmon Fis	hing Area 13			Salmon Fish	Ing Area 14A	1
		2SW		2SW		2SW		2SW			2SW	2SW	V	2SW	/	2SV	V	2SV	V	2	SW		2SW		2SW
	Returns		Spawne	rs	F	Returns		Spawners		Returns		Spawners		Returns		Spawners		Returns		Spawners		Returns		Spawners	
Year	Mi	in Max		Min	Max	Min	Max	Min	Max	Mi	n Ma	x Min	Max	Min	Max	Min	Max	Min	Max	Mi	n Max	Mi	in Max	x Mi	lin
labels	SFA9R2_L	[] SFA9R2_H[]	SFA9S2	_L[] SFA9S2	2_H[] S	SFA10R2_L SFA	10R2_H[SFA10S2_L SF	A10S2_H[SFA11R2	L SFA11R2_F	SFA11S2_L SI	FA11S2_H[SFA12R2_SI	A12R2_H	SFA12S2_S	FA12S2_F	SFA13R2_L[] S	FA13R2_H[]	SFA13S2_L[]	SFA13S2_H[]	SFA14AR2	L SFA14AR2_H	[SFA14AS2	_L SFA1/
1970) a	37 356	5	36	354	12	113	11	112	9	9 94	5 92	931	15	141	7	125	5 1300	3036	64	3 2050) <u> </u>	36 514	4 .	13
1971	3	32 305	1	30	301	18	174	17	171	8	0 76	3 74	750	9	85	7	82	2 1071	2850	65	3 2223	3 3	31 435	5	0
1972		22 214		22	213	11	107	11	106				905	18	174	16	170		3101				20 280		8
1973		13 406		41	402	35	336	31	327				906	13	121	10	116		3120				43 611		9
1974		57 230		57	228	47	186	45	184				681	25	101	24	99		2554				361		79
1975		38 352		87	351	20	80	19	78				838	24	96	22	92		4454				51 203		42
1976		5 261		64	258	20	117	28	115				660	12	48		47		3293						34
1977		34 136		33	134	35	138	34	113				513	12	70	12	61		2159				57 266		19
1978		26 103		24	99	31	138	27	117				312	7	20	5	25		3173				200		19
1978														14	29	13	54						23 93		19
1979				31 30	127 130	16 23	65 91	15	63				261		30	9	43		1117 3548				23 93 59 278		
													295	12	49										51
1981		33 330		77	320	58	233	55	228				1243	47	188	45	184		4471				9 436		95
1982		15 59		11	52	10	40	9	38				225	13	50		47		2298						99
1983		35 141		28	127	17	70	10	54				287	10	41	~ ~	39		3060						.63
1984		13 244		12	243	8	157	8	157				754		207	4	194		2226				27 518		18
1985		10 182	1	10	182	7	138	7	138				607	6	114	6	114		1505				20 378		20
1986		21 246		21	246	14	164	14	164				594	8	100	8	100		2190				45 539		44
1987	·	9 98	1	9	98	4	48	4	48	3	9 43	2 39	432	8	94	8	93		1813	60	5 1793		47 522		47
1988	1	12 150)	12	150	10	124	10	124	4	6 54	8 46	548	13	154	13	153	3 780	2350	76	4 2326	5 5	57 681	1 5	55
1989	1	16 164		16	164	11	112	11	112	3	6 37	0 36	370	7	70	7	70	339	939	33	4 931	1 3	39 400	ა ::	39
1990	2	26 200)	26	200	11	89	11	89	6	1 47	6 61	476	12	92	11	91	711	1851	69	8 1830) 6	59 541	1 (68
1991		8 64	1	8	64	3	26	3	26	2	7 21	1 27	214	9	73	9	73	684	1460	67	6 1448	3 5	52 406	δ :	51
1992	2 1	18 362	1	18	362	7	128	6	128	6	0 119	2 60	1192	17	335	16	333	3 1235	5357	119	7 5300	12	27 2505	5 1.	23
1993	4	40 312	1	40	312	20	153	19	152	8	6 67	3 86	672	21	162	21	162	2 1047	2848	101	8 2804	11	10 852	2 10	.06
1994	1	11 105	1	10	104	9	89	9	88	2	7 26	7 26	265	10	97	9	95	5 1390	3528	128	3 3366	5 4	47 466	6 4	44
1995	5 1	19 195	i	18	193	16	160	15	159	3	9 39	3 38	391	8	84	8	83	3 737	3058	64	3 2916	5 7	75 762	2 *	71
1996	5 2	27 184		26	183	36	253	35	250	8	4 58	0 82	576	22	154	22	152	2 1391	4279	128	0 4111	14	15 1004	4 1/	41
1997		23 172		22	171	32	240	31	238				784	28	215	28	214		5113				58 1193		55
1998		28 234		27	233	47	390	46	389			2 50	421	14	118	13	116		4248						41
1999		22 169		22	167	28	215	27	213				404	6	48	6	48		4643				19 1133		44
2000		36 212		35	210	35	206	32	200		9 46		463	12	71	12	71		6029				58 995		.64
2000		7 34		7	33	10	49	9	48		+0	6 21	403	7	37	7	36		2324				59 346		67
2001		5 34		5	33	10	66	10	48		5 16	0 25	155	6	41	6	41		2324				74 472		72
2002		5 33		5	33	8	57	8	56		1 21	· · · · ·	212	13	89	12	41		3011				74 472 39 612		85
2003		4 38		4	37	8	73	8	72		9 27		212	7	61	6	61		3255				59 612	-	61
		4 38 6 76		6	37 74	8	108	8	12	2				9	61 104	6	103						54 590 57 673		61 54
2005						27	108		106	2	0 24 5 30		241	17	104 97	17	103		4629						
2006				21	121			26	148				308						5376						38
2007		12 81		11	81	17	119	17	119	2			193	14	96	13	95		3912				97 682		96
2008		13 90		13	89	16	109	15	107	3	7 25		252	9	59	8	59		3451				35 586		82
2009		9 122		9	122	13	183	13	182	1	9 26		268	4	62	4	62		3858				54 752		52
2010		22 134		21	133	27	166	26	165		6 16		162	10	61	10	61		3951				04 643		01
2011		9 196		19	195	32	335	31	332		5 25		255	9	92	9	92		1498				6 895		84
2012		20 117		20	117	26	147	25	146				117	8	46	8	46		873						05
2013		1 194		11	193	19	328	18	326	1			299	5	83	4	82		3687		4 3621		3 567		32
2014		7 96		17	96	20	111	19	108				82	6	33	6	32		2482				60 332		59
2015	5 1	7 110		17	109	17	112	17	111	2	1 13	9 21	138	8	50	7	49	1219	4526	116	9 4446	5 11	5 745	5 11	13

Appendix 4.viii. (Continued). Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

Appendix 4.ix. Input data for small salmon returns to Québec by category of data used in the run-reconstruction.

	Small retu	rns								Small return	ns								
	Minimum									Maximum									
Year	C	:1	C2	C3	C4	C5	C6	FN Harvest	Other rivers		C	2 C3	6 C4	C5	C6	FN Harvest	Other rivers	FN	Commercia
Bugs labels	QCSmC1 L		mC2 L[]QCS	SmC3 L[]	QCSmC4 L[QCSmC5 L[]	QCSmC6 L[]	QCSmFn L[]	QCSmO L[]	QCSmC1 H	QCSmC2 H	QCSmC3 H	QCSmC4 H	QCSmC5 H[QCSmC6 H[QCSmFn H[QCSmO H[]	QCFnSm[]	QCCmSm[]
1970		0	0	0	C	0 0	0	0	0	Ċ) () (0 0	0	0	0	0		0
1971		0	0	0	C	0 0	0	0	0	C) () () 0	0	0	0	0		0
1972		0	0	0	C) 0	0	0	0	0) () () 0	0	0	0	0		0
1973		0	0	0	C) 0	0	0	0	0) () () 0	0	0	0	0		0
1974		0	0	0	C) 0	0	0	0	() () () 0	0	0	0	0		0
1975		0	0	0	C	0 0) (0		0	0		0
1976		0	0	0	0								0	0			0		0
1977		0	0	0	0									0		-	-		0
1978		0	0	0	0			-						0					0
1979		0	0	0	0		-		-	-				0	-	-			0
1980		0	0	0	0			-						0			0		0
1981		0	0	0	0	, .			-	-				0	0				0
1982		0	0	0	(-	-				-	0		-	0		0
1983		0	0	0	-		-		-				,						0
1983	383		5434	2955	460												0		-
1984			2271	2955															
	526				210														
1986			5193	2396	63									11198					
1987	1004		4775	3852	327			267				, , , , , , , , , , , , , , , , , , , ,							
1988	1119		5968	4404	468														
1989			4743	2924	301														
1990	1224		7332	4377	694														
1991	955		5851	3776	349			256											
1992			6928	4567	428														
1993			6325	3973	1029														7 362
1994	870		5928	3840	1051			408											
1995			3439	2697	1017														
1996	1501	10	1809	3600	477	4666	500			16122	1923	3 7370	797	7816	833	200	8	7	2 453
1997	1149	91	201	3457	292	3529	462	58	563	12089	242	2 7049	487	5882	770	97	938	3	5 353
1998	1128	35	1183	3578	328	5121	1127	58	0	11849	1406	5 7347	555	8536	1878	97	0	3	5 106
1999	1087	77	708	3194	1868	3 5401	1429	0	0	11556	5 74 ⁻	6536	3098	9002	2382	0	0	71	0 81
2000	1188	36	429	1116	602	2 7399	633	0	0	12635	5 458	3 2284	1004	14050	1055	0	0	82	1
2001	805	50	185	2632	266	3225	728		• 0	8588	3 228	5392	2 443	5374	1213	0	0	77	0
2002	1459	99	31	3189	689	4333	1448	0	0	15494	36	6530	1149	7222	2414	0	0	167	2
2003	1136	62	0	3203	721	3566	1512	0	0	11903	6 (6538	3 1201	5944	2520	0	0	97	2
2004	1374		107	6526	284				0	14177	127	/ 13104		8149		0	0		
2005	877		0	3689	794				0							0	0		
2006	1276		0	3736	1800			0				7584							
2007	851		0	3758	1710			0	0			7631			1707	0			
2008	1644		0	5542	2266			0				11261				0			
2009	887		0	3601	903			0	-			7306		3178		0			
2000			0	4801	993			0				9746				-	-		
2010	1799		0	5120	1365			0				10386				0			
2011			0	3615	584							7332				0			
2012			88	3185	411			0									0		
2013			00	3945	676			0	-			8003		4111		0	0		
2014			0	3945 5843	1396) 8003) 11853							

	Large return	IS							Large return	ns								
	Minimum								Maximum									
Year	C1							Other rivers		C2	C3	C4	C5	C6		Other rivers		Commercia
Bugs labels	QCLgC1_L[]	QCLgC2_L[]	QCLgC3_L[]	QCLgC4_L[]	QCLgC5_L[]	QCLgC6_L[]	QCLgFn_L[]	QCLgO_L[]	QCLgC1_H[]	QCLgC2_H[QCLgC3_H[]	QCLgC4_H[] QCLgC5_H[]	QCLgC6_H[]	QCLgFn_H[]	QCLgO_H[]	QCFnLg[]	QCCmLg[]
1970	0	0	0	0	0	0	C	0	0) () (0 0) () () C)	0
1971	0	0	0	0	0	0	C	C	0) (0 0)	0 0) () (0 0		0
1972	0	0	0	0	0	0	C	C	0) () ()	0 0) () () (0
1973	0	0	0	0	0	0	C	C	0) () (0 0) () () C		0
1974	0	0	0	0	0	0	C	C	0) () (0 0) () () C)	0
1975	0	0	0	0	0	0	C	C	0) () (0 0) () () ()	0
1976	0	0	0	0	0	0	C	C	0) (0 0) () () ()	0
1977	0	0	0	0	0	0	C	C	0) () (0 0) () () ()	0
1978	0	C	0	0	0	0	C	C	0) (0 0) () () ()	0
1979	0	0	0	0	0	0	C	C	0) (0 0) () () ()	0
1980	0	0	0	0	0	0	C	0	0		ó (0 0) () () ()	0
1981		-			0	0							0 0	-		-		0
1982	0	C	0	0	0	0	C	C	0				0 0) () () (0 0
1983	0	0	0	0	0	0	C	C			y ()	0 0) () () ()	0
1984	14119	9501	2922	3407	3712	5071	329	108	15631	978	6035	647	7 6187	7 8452	2 548	3 181	453	0 1305
1985	14015	7028	3836	345	9215	3351	329	76	15611	728	1 7809	57	7 15827	7 5586	548	3 127	362	3 16619
1986	18589	8598	6152	35	5877	4971	329	89	20602	8839	9 12596	6 6	1 9795	5 8284	1 548	3 149	451	9 2088
1987	17574	6715	5178	273	6335	3012	329	82	19017	6889	9 10575	5 45	8 10558	3 5019	9 548	3 137	446	6 2274
1988	21445	6432	7540	346	6789	4781	329	98	22979	6618	3 15336	57	6 11315	5 7969	548	3 164	474	7 1975
1989	20278	8503	5530	278	5718	4567	329	106	21906	8736	6 11252	2 46	5 9531	I 7611	548	3 176	290	5 18175
1990	17098	10803	8164	1365	5179	2424	442	112	18222	1104 ⁻	1 16613	3 227	6 8631	4040	737	7 187	290	0 16092
1991	19112	6988	7183	696	3856	357	242	101	20443	7192	2 14602	116	1 6427	7 595	403	3 168	433	5 16372
1992	18392	7360	7930	372	2687	1503	461	76	19578	7560	0 16149	62	2 4478	3 2505	769	127	455	0 1585
1993	14578	10133	2866	373	2649	333	423	52	15454	1146	3 5849	62	4 4414	1 555	5 705	5 87	397	6 11242
1994	16538	9172	2644	506	2853	145	427	60	17594	1024	1 5411	84	5 4755	5 242	2 712	2 100	449	6 1042
1995	21658	9598	1926	813	4390	154	246	31	22968	10936	3915	135	8 7317	256	6 410) 52	619	4 1003
1996	22679	5822	3843	577	2486	135	113	4	24117	694	1 7844	96	4 4155	5 225	5 189	9 7	611	3 7454
1997	18106	4221	2816	333	2865	138	48	9	19154	5154	1 5768	55	3 4775	5 229	80) 15	487	5 7202
1998	13180	4927	2861	347	2790	291	48	0	13891	5962	2 5907	59	2 4649	485	80	0 0	487	5 1038
1999	16912	842	2554	3661	3870	492	C	0	17700	995	5 5232	610	3 6450	838	3 C	0 0	368	3 47
2000	14568	619	3901	560	6420	563		0	15300	669	9 7947	93	3 10700	949) C	0 0	381	8 (
2001	17837	633	5320	241	3988	556		0	18889	879	9 10914	40	2 6647	926	6 C	0 0	357	4 (
2002	12335	8	4515	339	2103	345	Ő	0	13001	ę	9 9277		5 3505	5 575	5 C	0 0	316	4 (
2003	21853	0	5787	269		384	0	0	22893	. (11779	44	9 8148	641	C	0 0	354	1 (
2004	18369	107	4870	357	4432	401	0	0	19043	126	6 9170	59	5 7387	668	з с	0 0	355	8 (
2005	19154	0	3204	734	4815	351	0	0	20066	(0 6515	122	3 8025	5 585	5 C	0 0	306	2 (
2006	16704	0	3387	901	3945	403	0	0	17500	(6904	150	2 6575	672	2 0	0 0	351	2 (
2007	14832	0	3638	1301	3171	305	0	0	15604	. (7406	216	8 5285	5 508	s C	0	293	2 (
2008	15216	0	5187	1328	5423	390	0	0	16002	. (0 10595	221	3 9038	649) C	0	297	1 (
2009	18479	0	3727	950	4556	275	0	0	19412		7589	158	4 7594	458	S C	0	275	2
2010	21375	0	4488	1047	3656	338	C	C	22454		9157	174	4 6093	3 564	1 C	0 0	236	2
2011	26977	0	4697	1571	6007	483	C	C	28373	. (9529	261	9 10011	805	5 0	0 0	321	6
2012	17918	0	3665	904	4488	313	C	0	18837	. (7434	150	7 7481	522	2 0) C	302	3
2013	22026	205	4171	989	3938	339	C	0	23135	242	2 8461	164	8 6563	3 565	5 0) (289	5 (
2014	10954	0	2400	695	1593	1035	C	0	11504	. (4869	115	9 2655	5 1725	5 0) C	290	8 (
2015	18085	0	3995	1812	5048	568	C	0	19007	· (8104	296	5 8414	1 946	6 0) (276	5 (

Appendix 4.ix. (Continued). Input data for large salmon returns to Québec by category of data used in the run-reconstruction.

Appendix 4.ix. (Continued). Input data for small salmon spawners to Québec by category of data used in the run-reconstruction.

	Small spawn	ners					Small spawn	ers				
	Minimum						Maximum					
fear	C1	C2	C3	C4	C5	C6	C1	C2	C3	C4	C5	C6
Bugs labels	QCSSmC1_L	QCSSmC2_L[QCSSmC3_L	QCSSmC4_L	QCSSmC5_L	[QCSSmC6_L	QCSSmC1_H	QCSSmC2_H	QCSSmC3_H	QCSSmC4_H	QCSSmC5_H	QCSSmC6_H
1970	0	0	0	0	C	0 0	0	0	0	0	0	0
1971	0	0	0	0	C	0	0	0	0	0	0	0
1972	0	0	0	0	C	0 0	0	0	0	0	0	0
1973	0	0	0	0	C	0 0	0	0	0	0	0	C
1974	0	0	0	0	C	0 0	0	0	0	0	0	C
1975	0	0	0	0	C	0 0	0	0	0	0	0	C
1976	0	0	0	0	C	0 0	0	0	0	0	0	C
1977	0	0	0	0	C	0 0	0	0	0	0	0	C
1978	0	0	0	0	C	0 0	0	0	0	0	0	C
1979	0	0	0	0	C	0 0	0	. 0	0	0	0	C
1980	0	0	0	0	C	0 0	0	0	0	0	0	C
1981	0	0	0	0	C	0 0	0	0	0	0	0	C
1982	0	0	0	0	C) 0	0	0	0	0	0	C
1983	0	0	0	0	C	0	0	0	0	0	0	C
1984	3061	4342	1915	415	1264	5160	3316	4547	5013	747	2378	8599
1985	3960	1622	1025	209	4241				2844	351	8016	7307
1986	6337	3827	1499	63	5151	5133	7160	3955	3998	107	9630	8555
1987	7493			291	6411				6388	510		9168
1988	8173			419					7296	731	12059	
1989	7779			273					4894	475		
1990	8735				5437				7362	1068		4960
1991	7247		2465	316					6368	551	7166	
1992	5989		2937	370	3957				7667	657	7378	
1993	4852			747	3339				6626	1435		
1994	5506			894	3089			4197	6489	1596		729
1995	5348			877	2956				4534	1556		
1996	10636			372					6030	692		833
1997	8238								5842	461	5426	
1998	7734			289	4229				6116	516		
1999	8155								5837	2883		
2000	8291			519	5900				1861	921	12551	1005
2000	5329											
2001				658				36		1118		
2002	8180				2826					1141		
2004				278						468		
2004				716						1245		
2005					2372					2890		
2000	5768				1501					2651		
2007				1756						3266		
2000	6293				1633					1366		1759
2009										1576		
2010	12143									2027		
2011										861	5290	
2012	4959									602		
2013				328 560	2131					1011	3776	
2014				1109						2008		

	Large spaw	ners					Large spaw	ners				
	Minimum						Maximum					
Year	C1	C2	C3	C4	C5	C6	C1	C2	C3	C4	C5	C6
Bugs labels	QCSLgC1_L[QCSLgC2_L[QCSLgC3_L[QCSLgC4_L	[QCSLgC5_L[QCSLgC6_L[QCSLgC1_H	QCSLgC2_H	QCSLgC3_H	QCSLgC4_H	QCSLgC5_H	QCSLgC6_H
1970	0	0	0	C	0	0	0	0	0	0	0	
1971	0	0	0	C	0 0	0	0	0	0	0	0	(
1972	0	0	0	C	0 0	0	0	0	0	0	0	(
1973	0	0	0	C	0 0	0	0	0	0	0	0	
1974	0	0	0	C	0 0	0	0	0	0	0	0	
1975	0	0	0	C	0 0	0	0	0	0	0	0	
1976	0	0	0	C	0 0	0	G	0	0	0	0	
1977	0	0	0	C	0 0	0	0	0	0	0	0	(
1978	0	0	0	C	0 0	0	0	0	0	0	0	(
1979	0	0	0	C	0 0	0	0	0	0	0	0	(
1980		0	0	C			0	0			0	(
1981	0	0	0	C) 0	0	C	0	0	0	0	(
1982			-	-				-	0			
1983								0	0			
1984		7648		2357			11933		4974	5427		
1985		4991					11581		6098	572		
1986		5804					15672		10139		8416	
1987	13432	4791							8422		9053	
1988		4258		312					12177	542		
1989		6742					16273		8961	440		
1990		8463							13006			
1991	14061	5019					15392		11389	1061	5511	595
1991		4819							12711	575		
1992		6936							4792			
1993		5920							4460	787	4075	
1994		8323							3310			
1995		4417							6390		3593	
1990	13433	3393							4696			223
1997		4429							4895	520	4147	484
1998		747										
2000		570					14957 12669		4640 7368	5542 864		
2001	14527	505							8986			
2002		8		313					8833	539		
2003		0							11156			
2004		107		355			16232		8531	593		
2005		0		719			17397		6212			
2006		0							6572			
2007	12470	0							6971	2154		506
2008		0							10084	2151	7911	649
2009		0							7050	1483	6626	
2010		0							8595			564
2011	24130	0							9012			801
2012		0		868					6990	1471	6678	
2013		205		920					7991	1579	5550	564
2014	10089	0			1206				4607	1128		1430
2015	16672	0	3560	1749	4427	557	17594	0	7669	2902	7793	935

Appendix 4.ix. (Continued). Input data for large salmon spawners to Québec by category of data used in the run-reconstruction.

	i nese data are	e specific to the	e 1970 to 1983	period when d	letailed returns	by river categ	ory are not ava	liadie.
	Small returns		Large returns		Small spawne	ers	Large spawne	ers
Year	Min	Max	Min	Max	Min	Max	Min	Ma
Bugs labels	QCSm L[]	QCSm_H[]	QCLg_L[]	QCLg_H[]	QCSSm L[]	QCSSm_H[]	QCSLa []	QCSLg_H[]
1970	18904						0- 11	
1971	14969							
1972	12470							
1973	16585		68171	102256		16480		
1974	16791	25186	91455					
1975	18071	27106						
1976	19959	29938	77212					
1977	18190		91017					
1978	16971	25456						
1979	21683	32524	45197					
1980	29791	44686	107461	161192				
1981	41667	62501	84428					
1982	23699	35549	74870					
1983	17987	26981	61488			18045		
1984	0							
1985	0							
1986	0							
1987	0							
1988	0							
1989	0							
1990	0							
1990	0							
1992	0							
1993	0	0						
1993	0	0						
1995	0							
1996	0	0						
1997	0		0					
1998	0							
1999	0	0						
2000	0		0					
2000	0		0					
2001	0		0					
2002	0	0	0					
2003	0	0						
2004	0	0	0					
2005	0	0						
2000	0	0						
	0							
2008 2009	0							
2009	0							
2010	0							
	0							
2012								
2013	0							
2014	0							

Appendix 4.ix. (Continued). Year specific returns and spawner data for Québec for years when category splits are not available (1970 to 1983) used in the run-reconstruction.

	Returns of 2S	VV											
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return													Point
to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	estimate
										SF19_21R2_H			
Vinbugs labels	SF15R2_L[]			SF16R2_H[]		SF17R2_H[]	SF18R2_L[]	SF18R2_H[]	[]				
1970	8243			45798	31	60							
1971	3587			30669	29			2782					
1972	4980				402								
1973	6211			40492	206								
1974	7264			60090	386			6928					
1975	4353			39325	345					13797			
1976	7293			30758	575					10104			
1977	9174			73330	606				10872		16869		
1978	5458			26041	0				8272				
1979	1472			9306	459					4879			
1980	7102			48457	2				14094				
1981	4572			19268	40						11026		
1982	4314			41643	16	31	4135		4458	5353			
1983	3453			28419	17			5241	4134				
1984	3329	6092	12216	31455	13	26	2391	3573	1758	2854	15706	22627	
1985	4805	9500	14614	37625	8	15	921	4481	6894	12124	16541	23828	
1986	7831	15403	21617	55640	5	11	2274	11479	6755	11878	9891	14261	6176
1987	4836	9123	12524	32224	66			10422	3748	6591	6922	10043	
1988	7152	13998	14384	36938	96	185	2533	10205	4393	7735	4716	6697	3286
1989	4390	8492	9113	23385	149	287	2108	8600	4808	8469	6560	9437	3197
1990	4326	8369	14269	36639	284	545	1893	7684	3591	6320	5486	7918	
1991	2387	4668	14685	37736	188	361	2350	9628	2960	5213	7337	10563	2647
1992	4002	7787	21381	30728	95	183	2374	9577	2633	4634	6878	9809	2459
1993	1395	2684	15579	60246	22	43	1341	5317	2542	4470	4345	4820	2231
1994	3960	7745	13652	24887	169	310	1981	8094	1360	2396	3084	3495	1346
1995	2713	5333	25593	37215	384	576	1498	6160	2253	3969	3439	3998	1748
1996	3917	7754	11126	19117	394	591	3247	13507	3000	5278	4729	5397	2407
1997	2488	4898	8545	14244	387	581	3421	14254	1163	2045	2769	3176	1611
1998	1687	3260	6292	10783	385	577	2055	8560	924	1270	1372	1642	1526
1999	1780	3425	7098	11206	383	575	1557	6596	1419	1951	2375	2640	1168
2000	2270	4410	7560	11744	378	566	1467	6302	1078	1483	988	1206	
2001	3779	7442	14257	19289	376	564	1689	7251	1822	2506	1938	2279	788
2002	2335				372		1228		382				
2003	3947	7778	10991	16823	371	557	2380	10207	1854	2548	1056	1198	
2004	3005			18488	367	550							
2005	3422			19988	373	560	2217						
2006	2551	4973		17103	392		2114			1734			
2007	4267			15183	412								
2008	2848			11066	429				1793				
2009	3948			17076	402				827	1135			
2010	2978		7804	11581	439								
2011	7265			48573	653	980				2044			
2012	3230			15163	653	980		4118					
2012	5324			15865	993	1487	1871	8797					
2013	2704			11356	713		1266						
		5202	3034	11000	115	1007	1200	0044	415	503	172	217	

Appendix 4.x. Input data for 2SW salmon returns to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

Appendix 4.x. (continued). Input data for large salmon returns to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

	Returns o	flarge sal	mon										
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Point estimate
	SF15Lg_L	SF15Lg_	SF16Lg_L	SF16Lg_	SF17Lg_L	SF17Lg_	SF18Lg_L	SF18Lg_	SF19_21L	SF19_21L	SF23Lg_L	SF23Lg_	
Winbugs labels	[]	H[]	[]	H[]	[]	H[]	[]	H[]	g_L[]	g_H[]	[]	H[]	USALg[
1970	12681	16270	46462	49599	31	60	6161	7858	7273	9671	9691	13945	(
1971	5518	7102	28365	33409	29	29	2456	3198	5350	6773	8056	11573	653
1972	8441	16536	30146	45087	402	402	6095	6924	7460	9082	8890	12536	1383
1973	8393	16229	27771	42276	206	206	5376	6299	8049	10069	4760	6638	1427
1974	9950	19959	43249	66179	386	386	7119	7963	13138	15363	12187	16444	1394
1975	5510	10028	29826	45305	345	345	4483	4989	12261	13797	14829	20351	2331
1976	9596	18969	23943	36016	575	578	3578	4223	8873	10416	16128	22175	1317
1977	11053	21779	52673	77434	606	606	5175	6280	14119	16690	19165	26183	1998
1978	7277	14332	22653	30245	0	0	5954			12378			
1979	2886		9435	13507	459					6684			
1980	8768		37014	51008									
1981	9729		16708	28887	40	-	3234						
1982	7311	10700	26504	51475			5370			11390			
1983	5852		20309	35304				_					
1984	4214		12941	33321	13								
1985	7627	15080	16798	43247	8	-							
1986	10305	20267	25342	65228									6176
1987	7556		15734	40483									
1988	9933		17627	40483						12891			
1989	7701	14898	13955	35812									
1989	6362	14696	23164		284					12275			
1990	4773		23164	59479 62373									
1992	7411	14420	34573							6437			
1993	3487	6711	22602		22								
1994	6600			32992	169								
1995	4171	8199	30324		384				2532				
1996	6026			28035	394					6283			
1997	3828		14711	24521	387								
1998	2595			26060	385								1526
1999	2738			23026									
2000	3493												
2001	5815		22082	29875									797
2002	3592		11094	18077	372								
2003	6072		18783	28749		557							1199
2004	4623	9055	18589	32435	367	550	3427	13100	1302	1789	1402	1698	1316
2005	5265		17008	30057	373								994
2006	3924	7651	18805	32890	392	587	2746						1030
2007	6565	12957	16018	25734	412	618	1757	7037	701	959	689	841	958
2008	4382	8572	10377	19761	429	644	2623	10755	1928	2650	858	1105	1799
2009	6074	11970	17065	27543	402	602	1979	8335	1034	1418	1678	2158	2095
2010	4581	8972	15301	22708	439	658	2662	11005	1061	1451	1117	1398	1098
2011	11177	22223	24960	57144	653	980	4718	19229	1504	2065	2598	3421	3087
2012	4969	9750	11411	21661	653	980	1080	4733	788	1075			
2013	8190		10860	22992			2430						
2014	4160		8448	16849									
2015	5064		14152	23686									

Appendix 4.x. (continued). Input data for small salmon returns to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

	Returns o	fsmall sal	mon										
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return													Point
to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	estimate
	SF15Sm_	SF15Sm_	SF16Sm_	SF16Sm_	SF17Sm_	SF17Sm_	SF18Sm_	SF18Sm_	SF19_21Sm_	SF19_21Sm	SF23Sm_L[SF23Sm_H[
Ninbugs labels	L[]	H[]	L[]	H[]	L[]	H[]	L[]	H[]	L[]	_H[]]]	USASm[
1970	2834	6279	47779	67697	0	0	264	1073	16177	24106	5306	7521	C
1971	2113	4681	38388	54120	0	0	65	265	11911	18004	3248	4541	32
1972	2185	4699	48886	69270	0	0	131	530	11587	17992	1831	2506	18
1973	3010	6668	47190	66835	5	9	516	2095	14169	22159	5474	7012	23
1974	2226	4895	78091	110470	0	0	187	757	25032	39058	10195	12901	55
1975	2393	5298	69993	98443	0	0	112	454	10860	15753	18022	23101	84
1976	8667	14696	96504	136107	14	28	299			33009	22835	28864	186
1977	6085	12084	30621	42689	0	0	215	871	24599	37314	13738	16671	75
1978	4350	7749	29783	39927	0	0	78	316		10023	6271	7695	155
1979	4378	9495	50667	70714	2				24298	37514	15356	20517	250
1980	7994	15278	41687	58839	12	23	520	2108	34377	50250	25139	31483	818
1981	9380	17119	63278	108226	259	498	2797			48945	16826	21803	1130
1982	6541	13383	78072	133171	175	336	2150	8722	17619	27075	11811	15636	334
1983	2723	4638	24585	41332	17	32	212	858	9313	14068	9270	12592	295
1984	12003	15867	28714	49595	17	32	460	1867	18382	29867	15556	21678	598
1985	7003	15516	53393	92224	113	217	730	3167	24384	39541	13056	17928	392
1986	10813	23926	103230	178295	566	1088	965	3854	24369	39663	14274	20183	758
1987	9630	21220	74485	128644	1141	2194	1646	5713	27269	44266	13358	17662	1128
1988	13168	29092	107071	184904	1542	2963	1381	4833	24509	39750	16381	23084	992
1989	6357	13900	66069	114097	400	770	893	3208	25602	41557	17579	24521	1258
1990	7880	17314	73020	126115	1842	3539	983	3528	29471	48039	13820	19176	687
1991	4441	9828	53453	92327	1576	3028	1160	4166	9762	15955	13041	17685	310
1992	8853	19614	142416	204708	1873	3599	994	3531	13754	22269	13563	18404	1194
1993	5783	12812	70090	175096	1277	2454	1146	3892	13297	21681	7610	8828	466
1994	9136	20208	41773	59888	210	385	671	2425	3154	5393	5770	6610	436
1995	2902	6429	44357	63453	658	987	543	1985	8397	13873	8265	9458	213
1996	6034	13370	32067	45995	710	1065	2431	8958	13120	22293	12907	15256	651
1997	5797	12845	14377	24122	517	776	561	2134	3410	5863	4508	4979	365
1998	6288	13932	21965	32523	508	762	633	2419	8833	11927	9203	10801	403
1999	4936	10929	21494		413	620	705	2681	3971	5337	5508	6366	419
2000	7459	16520	31923		395	593	615	2428	6155	8312	4796	5453	270
2001	4947	10953	26496	36655	415	622	822	3205	2326	3138	2513	2862	266
2002	11719	25958	40432	54790	390	585	844	3319	5197	7015	3501	3991	450
2003	3119	6904	26530	39772	515	773	773	3088	2844	3837	2292	2716	237
2004	12091	26783	43242	62082	330	495	1092	4339	3847	5192	3454	4297	319
2005	4117	9116	28441	47190	343	514	781	3015	2870	3871	3597	4640	319
2006	8724	19322	30671	52560	331	497	869	3406	5144	6940	3720	4743	450
2007	4259	9430	23038	44016	275	413	718	2820	4198	5664	2466	3136	297
2008	13601	30129	25722	46587	298	447	1245	5061	7282	9831	5924	7691	814
2009	5169	11445			233	350	302		2066	2788	1603		241
2010	8187	18132			258						9114		525
2011	10234	22668		67884	291						4466		
2012	4350	9631	6914		291				346		178		
2013	4661	10320									894		78
2014	3697	8183			222				529	713	677		
2015	6228	13793								2648	1677		150

Appendix 4.x. (continued). Input data for 2SW salmon spawners to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

	Spawners	s of 2SW											
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Point estimate
	SF15S2	SF15S2	SF16S2	SF16S2	SF17S2	SF17S2	SF18S2	SF18S2	SF19 21	SF19 21	SF23S2	SF23S2	
Winbugs labels		H[]	L[]	H						S2_H[]			USAS2
1970	1156	3252	5346	8242	18					4234		4846	00/102
1971	510		6724	11354	0							6576	
1972	2367	6656	17031	31450	0					2865		9806	
1973	2873		19277	33170	0					3984		4412	
1974	3620	10183	31192	52012	0			790				12046	
1975	1769	4975	18536	31972	0							16209	194
1976	3530			22152	1							15583	
1977	4412		30623	54071	0					4539		18568	64
1978	2622				0					3455		8076	
1979	527	1482	3000	5806	3					2517		5650	
1980	3440		17667	30961	1							19005	426
1981	1380	3880	2392	10515	36		196			6439		7014	
1982	991	2786	8418		8		253			2053		7939	4643
1983	906	2547	5516		15			_		2800		3561	176
1984	2656	5402	11650	30889	13		-			2512		18798	
1985	4514		14019	37030	8					11990		18624	
1986	7279	14804	20606	54630	5					11748		11280	5570
1987	4122			31114	66							8597	278
1988	6582	13386	13801	36355	96			9954				5248	
1989	3944		8466	22739			2042			8396		9158	
1990	3886	7903	13669	36039	284					6260		7292	
1991	2193	4460	14200	37251	188		2275					9158	
1992	3639		20770		95					4589		8633	
1993	1239	2521	15239		22						3291	3654	
1994	3639	7401	13418		166							2680	
1995	2519	5124	25326		380							3652	
1996	3688	7502		18662	388		3166			5224		4585	240
1997	2316				385		3334					2565	161
1998	1512				382		2000			1261	1068	1302	
1999	1581	3217			379					1941	1934	2181	116
2000	2057				376					1477	805	1004	158
2001	3521	7161	13688	18674	374	564				2497	1699	2008	
2002	2120	4312	5332	8808	371						317	356	
2003	3683	7491		16372	368					2528		998	
2004	2770			17965	365					1401	1238	1492	
2005	3175	6457	10755	19354	371					890		914	
2006	2329		9336		390					1720		1023	
2007	3994	8124		14644	409					809		633	
2008	2618		5376	10584	429			9180		2450		953	
2009	3684	7494	10062	16500	401			7122		1118		1774	
2003	2743		7335	11078	438							877	148
2010	6902		20445	47555	652					2023		3196	
2012	2988		7603	14713	652					828		298	
2012	5019	10208	6909	15204	989					2834		526	
2013	2478		5606	11217	709							208	56
2014	3048		9518		709							309	1509

	Spawners of	large salmon											
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return										M			Point
to rivers	Min	Max	Min	Мах	Min	Max	Min	Max	Min	Max	Min	Max	estimate
Vinbugs labels	SF15SLa L[]	SF15SLg_H[]	SF16SLa L[]	SF16SLa H[]	SF17SLa L[]	SF17SLg_H[]	SF18SLa Líl	SF18SLa HI		_SF19_21SLg_ H[]	SF23SLa L[]	SF23SLg_H[]	USASLg[
1970	1779			8926							1451		
1971	785			12369							3888		
1972	4011										7246		
1973	3883										3050		
1974	4960										9090		
1975	2239										12335		
1976	4644			25940							11183		
1977	5315					0					13452		
1978	3496										5948		
1979	1033										4217		
1980	4248			32590							13190		
1981	2935								4907		3794		
1982	1679							1519			4903		
1983	1535							1264			92		
1984	3362			32721				1320			13675		
1985	7164							5010			13104		
1986	9577			64044			2811	12889			8004		
1987	6441							11711			6343		
1988	9141										3835		
1989	6919								6862		7099		
1990	5715				284						5576		
1991	4386										6833		
1992	6738										6511		
1993	3099										4026		
1994	6065			32682							2827		
1995	3873										3362		
1996	5674				388						4688		
1997	3563										2565		
1998	2326										1675		
1999	2433										2251		
2000	3165			24069							975		
2001	5417			28923							1831		
2002	3261										442		
2003	5666										919		
2004	4261							12895			1287		
2005	4884										791		
2006	3583										847		
2007	6145			24820							586		
2008	4028			18901					1912		767		
2008	5668										1565		
2009	4221										996		
2010	10619							18913			2532		
2011	3230			21019							300		
2012	7721										486		
2013	3812										220		
2014	3012			10042	409				410		220		

Appendix 4.x. (continued). Input data for large salmon spawners to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

Appendix 4.x. (continued). Input data for small salmon spawners to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run-reconstruction.

	Spawners of	small salmon											
	SFA 15		SFA 16		SFA 17		SFA 18		SFA 19-21		SFA 23		USA
Year of return													Point
to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	estimate
									SF19_21SSm	SF19_21SSm			
Winbugs labels						SF17SSm_H[]	SF18SSm_L[]		_L[]			SF23SSm_H[]	USASSm
1970	1417			45876									
1971	1056			38195									
1972	1034										C		
1973	1505		31703	51349								5575	
1974	1098								16680			10777	
1975	1195			78888	0				5819				
1976	2480			104130	8				14196				
1977	2467	7653	13270	25338					15120			5 12129	
1978	1398	4337		24833	0			248	2857			5680	
1979	2104		31829	51876					15716	28932	11640	16801	
1980	2996	9293	27791	44943	7			1655	18876	34749	19597	25941	
1981	3183	9874	35423	80370	151	390	1762	8908	21096	38837	7805	5 12782	100
1982	3038	9027	51324	106423	102	263	3 1354	6847	11244	20700	6532	2 10357	29
1983	820	2486	13298	30045	10	25	5 133	674	5653	10408	5132	8454	25
1984	1620	4971	7389	28271	10	25			13658	25143	10290	16412	
1985	3557	10936	32275	71106	66	170	145	1788	18024	33181	8164	13036	
1986	5589	16990	71918	146983	330	852	63	1729	18187	33481	10725	5 16634	66
1987	4867	14920	49971	104131	665	1718	527	3075	20213	37210	10257	14561	108
1988	6664	20468	71967	149800	899	2320	344	2388	18125	33366	13061	19764	92
1989	3191	9741	37696	85724	233	603	3 232	1650	18973	34928	13124	20066	108
1990	3996	12190	46902	99996	1074	2771	229	1750	22080	40648	10025	5 15381	61
1991	2215	6872	39648	78522	919	2371	271	2068	7363	13556	9495	5 14139	23
1992	4426	13728	116657	178949	1092	2818	3 189	1634	10125	18640	9485	5 14326	
1993	2891	8968	52050	157056	745	1922	2 261	1805	9970	18354	5762	6868	44
1994	4554	14125	25649	43764	118	292	2 179	1266	2661	4900	4965	5 5738	42
1995	1451	4501	34650	53746	250	375	5 148	1055	6512	11988	8025	5 9218	21
1996	3017	9359	19511	29260	258	387	1005	5596	10909	20082	11576	3 13892	65
1997	2899	8991	8702	15524	256	384	203	1290	2917	5370	3971	4433	36
1998	3144	9752	13997	21387	255	382	2 228	1464	8818	11912	8775	5 10348	40
1999	2465	7646	12193	17943	253	380	347	1837	3895	5261	5196	6048	41
2000	3727	11560	18837	26196	252	378	314	1717	6148	8305	4455	5 5087	27
2001	2470	7663	15703	22815	250	376	6 403	2217	2315	3127	2210	2530	26
2002	5857	18166	25458	35509	249	373	3 426	2334	5180	6998	3232	3689	45
2003	1557	4829	15727	24997	248	371	396	2201	2829	3822	2069	2469	23
2004	6043	18744	27425	40613	246	369	496	2934	3833	5178	3229	4039	31
2005	2056								2854				
2006	4359			35085		370	358		5119				
2007	2127		14420	29105	248								
2008	6798			30904				3189			5729		
2009	2581	8007	5867	13313									
2010	4090												
2011	5114			45125									
2012	2172			7895	290				343				
2013	2328		5360	12437	272				919		870		
2014	1845		3943		220				527		669		
2015	6036				229				1960		1654		

Appendix 5: Model walkthroughs

Summaries of the data preparation, model running and output processing were presented at a one-day workshop prior to the 2014 WGNAS meeting and provided step by step walkthroughs of the assessment processes. Where appropriate these have been updated in 2017.

NEAC pre-fishery abundance and national conservation limit model in R

[NB: Instructions apply to model version: "NEAC_PFA_CL_RR_model_2015-varM_v12" as used in 2015]

1) Introduction

This program performs the run-reconstruction estimation of pre-fishery abundance (PFA) of maturing and non-maturing 1SW salmon for each country (and region) in the NASCO-NEAC area. PFA is estimated for January 1st in the first sea winter. The program also establishes the pseudo stock–recruitment (S–R) relationship between lagged egg deposition and Total 1SW PFA, and applies a hockey-stick S–R analysis to estimate the National/Regional Conservation Limit where river-specific CLs are not available. The original model is described by Potter *et al.* (2004); minor changes to the estimation approach used for different countries and regions have been reported in the annual reports of WGNAS.

The model also estimates the proportion (mean and SE) of the Faroes catch originating in different countries/regions based on the genetic analysis of scales collected in the fishery between 1993 and 1995 and the estimated PFA for each country/region since 2001, when no fishery has been operating at Faroes. This requires the model to be run once to provide the catch proportions and then run a second time to provide a full PFA assessment. The catch proportions are also used in the Catch Options model.

2) To get started

2.1) Load RStudio or R;

2.2) Set up a folder from which you will run the program;

2.3) Use folder and file names without spaces;

2.4) Put the program, the input files (annual and multiannual) and the summary data file (see 6f) in this folder.

3) Input Data

- 3.1) Annual data (filenames: Annual-data-XX-YY.txt)
 - 3.1.1) There is a file for each country (XX) and region (YY) which contains the 40+ year time-series of data on catches, exploitation rates and non-reporting rates (plus additional data for some countries).
 - 3.1.2) To read the .txt files, it is easiest to open them from within Excel. i.e.
 - Open Excel;
 - select the correct folder;
 - click on 'Open'
 - You will probably need to change the setting in the lower right corner of the open box from 'Excel files' to 'All files';

- Double-click on the file you want to open and it should open the 'Text Import Wizard';
- select 'finish' (If this doesn't work reopen the file, but select 'Delimited' at step 1, 'Tab' at step 2 and 'General' at step 3.)
- 3.1.3) Do not add any formatting to the file. If loading a new version of a file that has been saved in Excel (e.g. after addition of a new year's data), re-save the file by clicking 'Save As' and selecting 'Text (Tab delimited)' from the 'Save as type' list. This will remove the formatting and add the .txt extension.
- 3.1.4) Do not change the file name.
- 3.1.5) Close and save the file before running the programme. You will be prompted to confirm that you want to lose the formatting; click 'yes'.
- 3.2) Multiannual-data (file-name: 'Multiannual-data.txt' or similar)
 - 3.2.1) This file contains most of the other parameters used in the model including: smolt age composition, fecundity and sex ratios by region, M, etc.
 - 3.2.2) The second value listed is the 'lastdatayear' which needs to be updated to the latest year for which data are provided in the Annual-data-XX files.
 - 3.2.3) The file is not formatted in columns so can be read easily in Notebook, which should be selected automatically if you click on the file to open it. (NB: If you open the file in Excel, don't save it because it will probably add "" marks to each line.
 - 3.2.4) All blank lines and lines starting with '#' are ignored in this file. Apart from these:
 - The first line must start with 'list(
 - The last line must be ')'
 - All other lines must be 'variable name' <- number, followed by a comer (except for the last data line which has no comer).
 - 3.2.5) If the module estimating the composition of the Faroes catch is run (see below) the new values must be inserted at the end of the multiannual data file in place of the current ones.
 - 3.2.6) Save the file before running the model.
- 4) Model structure
 - 4.1) Introductory section: contains working directory, source files and various parameters controlling the way the program runs (some of these will need to be changed for your laptop (see below).
 - 4.2) Functions: functions are sections of code that the program calls up to repeat the same job. They have to be run before they are first called by the program; this is achieved by placing them at the beginning of the code. The main functions run the hockey-stick analysis for the NCL model and output certain figures and tables.
 - 4.3) Faroes and Greenland sections: these sections calculate the harvest in the distant water fisheries.

- 4.4) NEAC country/regions sections: there is a section for each country (in alphabetic order) and region to calculate the main outputs of the R-R model.
- 4.5) Output summaries: this section creates NEAC summary figures and tables and the country/region data files for the Winbugs Forecast Model.
- 4.6) Faroes catch composition: The final section estimates the proportion (mean and SE) of the Faroes catch originating in different countries/regions. This requires the model to be run once to provide the catch proportions and then run a second time to provide a full PFA assessment.
- 5) Running the code from RStudio
 - 5.1) Open R Studio
 - 5.2) Select "File/Open File" and use the browser to select and open the code file; the code should open in the Top left panel. The code will have a name like "NEAC_PFA_CL_RR_model_2015_xxx"
 - 5.3) If you have been using the code recently, you can select "File/Recent Files" and select the file from the drop-down list (if it is there); you can open several code files simultaneously and they appear as tabs above the Top Left panel.
 - 5.4) To set up the code for your PC/laptop, R-click on the code and scroll down to:

line 40 -enter the full path name of the working directory (replace the text between the parentheses with the full pathname of the folder containing the code on your laptop (e.g. "D:/Modelling_NEAC/PFA_NCL_R/2014").

line 45 -ensure that the text between the parentheses shows the correct filename for the multi-annual data file.

lines 77–86 -select which countries you wish to run the assessment for by setting "run-XX": 1 = run country XX; 0 = do not run. The summaries will only be run if all countries are set to 1.

line 82 -set "PrintFigs" equal to '1' to output the summary figures (or any other value not to output them); otherwise "0".

line 89 -set "WinbugsFiles" equal to '1' to output the data files for the Bayesian forecast model (or any other value not to output them) ; otherwise "0"..

line 92 -set "PrintCountryTables" equal to '1' to output summary output data for each region that is run (or any other value not to output them); otherwise "0"..

line 98 -set "RunFaroeseCatchSplitEstimation" equal to "TRUE" to run the estimation of the Faroes catch composition; otherwise "FALSE".

5.5) You do not need to save your changes before you run the code. [If you wish to save any changes, use "File/Save" or "File/Save As" as normal. It's a good idea to include the extension ".R". NB: You will be prompted to save the file before you close it.]

5.5.1) To clear the 'console' area (lower left panel) press "Ctrl-L"

- 5.5.2) To run the program press "Ctrl-Alt-R"
- 5.5.3) You will see when part of the code run in console area. Errors will show in red. The run is complete when the final line shows ">"
- 6) Running the program from R
 - 6.1) Open R Studio
 - 6.2) Select "File/Open script" and use the browser to select and open the code file; the code should open in a separate panel. The code is currently called "NEAC_PFA_CL_RR_model_2014"
 - 6.3) To set up the code for your PC/laptop, R-click on the code and scroll down to:

SET WORKING DIRECTORY (wd): In line starting "wd <-" replace the text between the parentheses with the full pathname of the folder containing the code on your laptop (e.g. "D:/Modelling_NEAC/PFA_NCL_R/2014").

SET "run_XX": in the lines starting "run_XX <-" select which countries you wish to run the assessment for by setting "run-XX": 1 = run country XX; 0 = do not run. The summaries will only be run if all countries are set to 1.

SET 'PrintFigs': set "PrintFigs" equal to '1' to output the summary figures (or any other value not to output them).

SET 'WinbugsFiles: set 'WinbugsFiles'' equal to '1' to output the data files for the Bayesian forecast model (or any other value not to output them).

SET 'PrintCountryTables': set "PrintCountryTables" equal to '1' to output summary output data for each region that is run (or any other value not to output them).

SET /RunFaroeseCatchSplitEstimation': set "RunFaroeseCatch-SplitEstimation" equal to "TRUE" to run the estimation of the Faroes catch composition; otherwise "FALSE" [SEE BELOW]

- 6.4) You do not need to save your changes before you run the code, but you may wish to save a version to be safe. To do this use "File/Save" or "File/Save As" as normal. It's a good idea to include the extension ".R". NB: You will be prompted to save the file before you close it.
- 6.5) To run the program select "Edit/run all"
- 6.6) You will see when the code runs in the 'R console' panel. Errors will show in red. The run is complete when the final line shows ">"
- 7) Running the Faroes stock composition
 - 7.1) The 'Multiannual-data' file contains the latest estimates of the composition of the Faroes catch by European country/regions based on the results of the genetics analysis reported in the 2015 WGNAS report and the 2001-14 PFA outputs. These estimates may be updated if new genetics data are provided or additional years of PFA estimates are to be included.

- 7.2) To run the estimation, ensure that all "Annual-data-XX" files have been updated.
- 7.3) SET 'RunFaroeseCatchSplitEstimation' to "TRUE" and run the model.
- 7.4) The new stock composition parameters will be output in the file "Faroes_split_estimate.txt"; these data should be copied into the end of the "Multiannual-data" file to replace the values already there. These data are also required for the Catch Options models.
- 7.5) Reset SET 'RunFaroeseCatchSplitEstimation' to "FALSE" and run the model again to produce full updates of the PFA estimates.
- 8) Output files

The program produces the following outputs (if requested):

8.1) National plots: (filenames "Figure-XX")

PDF files showing the national plots currently used in the WG report. This includes: maturing and non-maturing 1SW PFA; returns and spawners for 1SW and MSW; homewater exploitation rates; and total catches (incl. non-reported) for each country (XX). It also shows the pseudo stock-recruitment hockey-stock plots for each region; these show the estimated CL, where this is used in the assessment.

8.2) Regional data: (filenames "Region_data_XX_YY")

Excel files showing PFA, returns, catch, exploitation rates and spawners for 1SW and MSW fish and total eggs and lagged egg estimates for each country (XX) and region (YY).

8.3) Input files for Forecast analysis: (filenames: "Winbugs_Data_XX_YY")

Excel files for each country/region containing mean and s.d. estimates for the simulations for lagged eggs, 1SW returns and MSW returns.

8.4) Summary tables by country:

- Median spawner numbers
- Conservation limits and SERs
- Maturing 1SW PFA
- 1SW returns
- 1SW spawners
- Non-maturing 1SW PFA
- MSW returns
- MSW spawners
- 8.5) Summary plot for N-NEAC and S-NEAC
- 8.6) A formatted Excel workbook is set up to link to the output files and format the tables ready for use in the WGNAS report.
- 9) Common problems
 - 9.1) The code will crash if an output file (Figure or Table) is left open. The error message (in red) may say:
 - Error:: cannot open file 'Fig-XX'

or

- Error in: : cannot open file 'Region_data_XX.csv': Permission denied
- 9.2) It doesn't matter if an input file is open, but the program may not read the latest version if it has not been saved.

More problems to be added when they are found!

