## Stock Annex: Salmon (Salmo salar) in Northeast Atlantic

Stock-specific documentation of standard assessment procedures used by ICES.

| Stock | Atlantic salmon; sal.27.neac_SA |
| :--- | :--- |
| Working Group | Working Group on North Atlantic Salmon (WGNAS) |
| Created | 28 March 2014 |

Last updated
Last updated by

April 2019
WGNAS meeting, 2019

## 1 General

### 1.1 Stock definition

### 1.1.1 Background

Atlantic salmon, Salmo salar L., have a wide range of life-history strategies. Most forms are anadromous, however, with a juvenile phase in freshwater followed by a period at sea feeding and growing, during which the fish undergo extensive migrations in the open ocean, before they migrate back to freshwater to breed. Most Atlantic salmon return to their river of origin to spawn. This precise homing behaviour has resulted in groups of fish originating in different rivers or tributaries becoming genetically distinct as they adapt to the particular conditions that they face in their home river and along their migration routes. As a result, fish from one river or tributary can differ from fish originating in other rivers/tributaries which have become adapted to a different set of conditions. These subgroups comprise genetically distinct 'populations' and these are regarded as the basic biological units of the Atlantic salmon species.

Large rivers and their tributaries can support several, genetically distinct populations, each with separate spawning areas within the main-stem of the river or its tributaries. In most instances, however, it is not possible to demarcate clear population boundaries within a river, and managing stocks and fisheries at this level of detail would be very complex. Thus, while there is a need to protect the sustainability of these units, the primary management unit (e.g. for reporting catch statistics and regulating fishing) is enerally taken to be the river stock, comprising all fish originating in eggs laid within theriver.

Atlantic salmon are native to the temperate and subarctic regions of the North Atlantic Ocean and there are over 2000 rivers draining into the North Atlantic that support the fish, about 1500 of which discharge into the Northeast Atlantic. In this area, salmon distribution extends from northern Portugal to northern Russia and Iceland, while in the Northwest Atlantic, the species ranges from northeastern USA (Connecticut) to northern Canada (Ungava Bay).

Ideally, the management of all individual river stocks, and the fisheries that exploit them, might be based upon the status of each individual population. This is not always practical, however, particularly where decisions relate to the management of distant
water salmon fisheries, which exploit large numbers of stocks originating in broad geographic areas. WGNAS has therefore had to consider how populations or river stocks should be grouped in providing management advice. For this purpose, groups have been established which fall within the meaning of a stock as 'an exploited or managed unit' (Royce, 1984) and that are consistent with the ICES (1996) definition of salmon 'stocks' as 'units of a size (encompassing one or more populations) which provide a practical basis for the fishery manager'. The issues around the grouping of Atlantic salmon stocks for the provision of management advice are reviewed in detail in Crozier et al. (2003). Such stock groupings have typically been referred to as stock complexes.

Salmon mature at various sea ages, typically returning to freshwater to spawn after one to three years at sea, but also sometimes at older sea ages; this varies widely between populations. Those salmon that return after one year at sea are referred to as one-sea-winter (1SW) salmon, or grilse, with older fish categorised as $2 \mathrm{SW}, 3 \mathrm{SW}$, etc. In practice, however, for management purposes these older sea age fishare typically aggregated and collectively referred to as multi-sea-winter (MSW) salmon. The sea age when salmon become sexually mature depends on genetics as well as growing conditions in the sea, and possibly freshwater, although the precise proximate factors initiating homeward migration are unknown (Hansen and Quinn, 1998). The sea age of Atlantic salmon is important in the context of stock definition since these different groups of fish have different migration routes, return at different times and are differentially exploited in fisheries. Thus, for example, it is only potential MSW salmon that are exploited in the distant water salmon fishery that operates off the west coast of Greenland.

### 1.1.2 Management framework for salmon in the North Atlantic

The advice generated by ICES is in response to Terms of Reference posed by the North Atlantic Salmon Conservation Organisation (NASCO), pursuant to its role in international management of salmon. NASCO was set up in 1984 by international convention (the Convention for the Conservation of Salmon in the North Atlantic Ocean), with a responsibility for the conservation, restoration, enhancement, and rational management of wild salmon in the North Atlantic. NASCO now has six Parties that are signatories to the Convention, Canada, Denmark (in respect of Faroe Islands and Greenland), the EU (which represents its Member States), Norway, Russia and the USA. While sovereign states retain their role in the regulation of salmon fisheries for salmon originating in their own rivers, fisheries within the jurisdiction of one Party that exploit salmon originating in the rivers of another Party may be regulated by NASCO under the terms of the Convention. This is currently the case for the distant water salmon fisheries at Greenland and Faroes.

NASCO discharges these responsibilities via three Commission areas shown below:


While homewater fisheries are not regulated directly by ASCO, national/ regional jurisdictions seek to comply with NASCO agreements and guidelines in exercising their responsibilities. In particular, NASCO's Agreement on the Adoption of a Precautionary Approach states that an objective for the management of salmon fisheries is to maintain the diversity and abundance of salmon stocks, and NASCO's Standing Committee on the Precautionary Approach interpreted this as being "to maintain both the productive capacity and diversity of salmon stocks" (NASCO, 1998).

NASCO's Action Plan for Application of the Precautionary Approach (NASCO, 1999) provides an interpretation of how this is to be achieved:

- "Management measures should be aimed at maintaining all stocks above their conservation limits by the use of management targets".
Socio-economic factors could be taken into account in applying the precautionary approach to fisheries management issues".
"The precautionary approach is an integrated approach that requires, inter alia, that stock rebuilding programmes (including as appropriate, habitat improvements, stock enhancement, and fishery management actions) be developed for stocks that are below conservation limits".

In requesting scientific advice from ICES, NASCO asks for an annual review of events in the salmon fisheries and of the status of salmon stocks around the North Atlantic; NASCO also requests management advice for stocks in each of the Commission Areas. In fulfilling these requirements, three specific purposes have been identified for which stock groupings may be required (Crozier et al., 2003):

- providing descriptions of the status of stocks;
- developing models to estimate and/or forecast pre-fishery abundance (PFA); and
- developing management advice for the distant water fisheries.

Crozier et al. (2003) further noted that there is no reason to assume that the same stock groupings should be used for all these purposes, indeed both the criteria used (e.g. geographical or biological features) and the resulting groups are likely to differ.

### 1.1.3 Stock groupings used by WGNAS in providing management advice

As noted above, Atlantic salmon would, ideally, be assessed and managed on the basis of river-specific stock units. In reality, $<25 \%$ of the rivers with salmon populations in the North Atlantic are so assessed (Chaput, 2012; ICES, 2013). Consequently, stock status is often, of necessity, assessed at broader regional, national and subcontinental scales. While there might be merit in grouping stocks according to biological criteria (which could cross jurisdictional boundaries), it has generally been considered that the difficulties of collecting data in a similar format in different jurisdictions is likely to outweigh the benefits of using such groups (Crozier et al., 2003). It is also recognised that compilations of data on stocks within each jurisdiction are of importance to regional / national managers. As such, regional / national stock groups are typically used by ICES in providing advice on the status of stocks, with additional information compiled on biological groups (e.g. sea ages) as required.

ICES has previously provided information on the status of stocks in the Northeast Atlantic Commission (NEAC) area by region or by country (as well as sea age). For the North American Commission (NAC) area similar information is provided for the USA and the five main provincial regions in eastern Canada: Labrador, Newfoundland, Québec, Gulf and Scotia-Fundy.

In providing management advice for the mixed-stock distant-water fisheries, broader scale stock groupings have been considered appropriate. For the NAC area this is based on the six geographic regions of North America detailed above. For the NEAC area, the following national groupings have been used in recent years to provide NASCO with catch advice or alternative management advice for the distant-water fisheries at West Greenland and Faroes.


Iceland (south/west regions) ${ }^{1}$
${ }^{1}$ The Iceland stock complex was split into two groups for stock assessment purposes in 2005 (ICES, 2005), largely on the basis of tag-recapture information. Prior to 2005, all regions of Iceland were considered part of the Northern NEAC stock complex.

These groups were deemed appropriate by WGNAS as they fulfilled an agreed set of criteria for defining stock groups for the provision of management advice that were considered in detail at the 2002 WGNAS meeting (ICES, 2002) and re-evaluated at the 2005 WGNAS meeting (ICES, 2005). ICES subsequently noted, however, that provision of catch advice for NEAC stocks in the distant water fisheries should preferably be
based on a larger number of smaller management units, similar to those used in the NAC area (ICES, 2010a; 2011). Such an approach was developed at the 2013 WGNAS meeting (ICES, 2013) and indicative catch advice was provided at the country level as well as the Southern and Northern NEAC stock complexes. ICES (2017) is awaiting feedback from NASCO on the choice of management units.

Salmon from most NEAC stocks mix in the Norwegian Sea in autumn and winter, and were exploited by the fishery at Faroes (the Faroes fishery has not taken salmon since 2000). While there is evidence that some salmon from NAC rivers have been caught in the Norwegian Sea, they are currently not considered in the NEAC assessments, although this is now under review. Recent genetic information suggests that more North American fish than previously thought were exploited in the fishery at Faroes. Further details on the results of these investigations are provided in Section 3.3.3 of this report and potential options tor accounting for these fish in future catch advice is provided in Section 3.6. To date, consideration of the leyel of exploitation of national stocks in the Faroes fishery (ICES, 2005) resulted in the proposal that catch advice for the fishery should be based upon all NEAC area stoeks and both 1SW and MSW fish.

In contrast, the fishery to the west of Greenland operates in an area where salmon from all North America and some Northeast Atlantic stocks mix in their second summer at sea. Catch advice for this fishery is thus based on non-maturing (potential MSW) fish from all regions of North America, while consideration of the level of exploitation of national stocks in the fishery from NEAC, resulted in catch advice being based upon only Southern NEAC non-maturing 1SW (potential MSW) fish (ICES, 2005).

### 1.2 Fisheries

Most exploitation of Attantic salmon is restricted to fisheries close to or within the rivers of origin of the stocks; these bomewater fisheries take adult fish that are mainly returning to these rivers to spawn. As noted above, these fisheries are not directly regulated by NASCO since the Parties retain responsibility for the regulation of fisheries for satmon originating in their own rivers. However, NASCO can regulate fisheries undertaken by a Party that take salmon originating in another Party's rivers, such as is the case for the distant-water fisheries at Greenland and Faroes. These fisheries take salmon originating in a large number of rivers over a wide geographical


### 1.2.1 The Northern Norwegian Sea Fishery

A longline fishery for salmon in parts of the Norwegian Sea, north of latitude $67^{\circ} \mathrm{N}$, commenced in the early 1960s. Several countries participated in this fishery and the pattern of fishing, area of operation and catches changed markedly over the years. At its peak in 1970 this fishery harvested almost 1000 tonnes of salmon.

The Convention for the Conservation of Salmon in the North Atlantic Ocean, which resulted in the formation of NASCO, came into force in October 1983. The Convention created a large protected zone, free of targeted fisheries for Atlantic salmon in most areas beyond 12 nautical miles from the coast. An immediate effect was the cessation of the salmon fishery in the Northern Norwegian Sea outside the Faroes EEZ, with the last catches in this area reported in 1984 (ICES, 2013).

In the late 1980s and early 1990s, NASCO acted through diplomatic initiatives to address fishing for salmon in international waters by vessels registered to non-NASCO Parties. There have been no reports of such activities since this period.

### 1.2.2 The Faroes fishery

The fishery in the Faroes area commenced in 1968 with a small number of vessels fishing up to 70 miles north of the Faroes; initially catches increased slowly up to 40 tonnes in 1977. Danish vessels participated in the fishery between 1978 and 1982 and, at the same time, catches started to increase rapidly, peaking at 1025 tonnes in 1981. Several factors contributed to this increase: the season was extended, more vessels entered the fishery, and the fishery shifted northwards.

From 1982, the Faroese Government agreed to a voluntary quota system, involving a total catch of 750 tonnes in 1982 and 625 tonnes in 1983 ( 255 boats allowed 25 tonnes each). Since NASCO's establishment, regulatory measures or decisions have been agreed by the Northeast Atlantic Commission in most years (Table 1.2.2.1), These have resulted in greatly reduced allowable catches in the Faroese fishery, reflecting declining abundance of the salmon stocks. There has been no commercial salmon fishery targeting salmon around the Faroes since the early 1990s. Catches in the fishery are presented in Figure 1.2.2.1.


Figure 1.2.2.1. Nominal catch of salmon (tonnes, round fresh weight) in the Faroese longline fishery, 1960-2016.

Table 1.2.2.1. Summary of Regulatory Measures agreed by NASCO for the Faroese Salmon Fishery (courtesy of NASCO).

| Year | Allowable catch (tonnes) | Comments/other details in the measures/decisions |
| :---: | :---: | :---: |
| 1984-85 | 625 |  |
| 1986 | - |  |
| 1987-89 | 1790 | Catch in any year not to exceed annual average (597t) by more than $5 \%$. |
| 1990-91 | 1100 | Catch in any year not to exceed annual average (550t) by more than $15 \%$. |
| 1992 | 550 |  |
| 1993 | 550 |  |
| 1994 | 550 |  |
| 1995 | 550 |  |
| 1996 | 470 | No more than 390 tonnes of the quota to be allocated if fishing licences issued. |
| 1997 | 425 | No more than 360 tonnes of the quota to be allocated if fishing licences issued. |
| 1998 | 380 | No more than 330 tonnes of the quota to be allocated if fishing licences issued. |
| 1999 |  | No more than 290 tonnes of the quota to be allocated if fishing licences issued. |
|  |  | No more than 260 tonnes of the quota to be allocated if fishing licences issued. |
|  | No quota set | It is the intention of the Faroese authorities to manage the fishery in a precautionary manner with a view to sustainability, and to make management decisions with due consideration to the advice from ICES concerning status of stocks contributing to the fishery. |
|  | No quota set | It is the intention of the Faroese authorities to manage the fishery on the basis of the advice from ICES concerning status of stocks contributing to the fishery in a precautionary manner with a view to sustainability and taking into account relevant factors such as socio-economic needs and other fisheries on mixed stocks. |
| 2007 | No quota set | It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs. |
| 2008 | No quota set | It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES |



Note: The quotas for the Faroe Islands detailed above for the period 1984-2000 were agreed as part of effort limitation programmes (limiting the number of licences, season length and maximum number of
boat fishing days) together with measures to minimise the capture of fish less than 60 cm in length. The measure for 1984/85 did not set limits on the number of licences or the number of boat fishing days.

The Faroes salmon fishery operated from November through to May. The salmon caught in the fishery originated almost entirely from European countries with salmon from many countries being present in the area (Jacobsen et al., 2001). Small numbers of tagged fish originating in North America were also recaptured in the fishery (e.g. ICES, 1991), but excluded from catch advice. Genetic investigations, based on salmon scales removed from fish caught in the fishery in the 1980s and 1990s, suggested North American fish may make a larger contribution to the Faroes fishery than originally indicated (ICES, 2015). There was no consistent seasonal trend in the estimated proportion of North American fish in the catches at Faroes and so the overall percentages for 1SW $(5.7 \%)$ and MSW ( $20.5 \%$ ) salmon have been used in subsequent analyses. (ICES, 2015). WGNAS has been asked to consider the implications of the findings in providing future catch advice to NASCO.

The fishery exploited mainly 2SW fish, although some 1SW and 3SW fish were also caught. Small salmon ( $<60 \mathrm{~cm}$ total length) in their first winter at sea were required to be discarded. Large numbers of farmed salmon were also observed at Faroes and there is evidence that farmed salmon escaping from net pens in Norway entered this area (Hansen et al., 1987; Hansen and Jacobsen, 2003). Such farmed fish accounted for a significant proportion of the catch; in the early 1990s, the proportion of farmed fish in this area was estimated at between 25 and 40\% (Hansen et al., 1999).

Tagging studies (of adułt fish caught in the fishery) have indicated that some fish caught at Faroes were apparently on theit way westwards, as they were reported from West Greenland later the same year (Jákupsstovu, 1988). However, salmon tagged at West Greenland were also reported in the area north of the Faroes the following year (ICES, 1984). Thus, salmon of European origin are believed to move through the Faroese area on their way to the feeding areas in the West Atlantic as well as on their return to homewaters.

Limited fishing at West Greenland is reported as far back as the early 1900s, although the present fishery dates from 1959 when local fishermen began setting fixed gillnets from small boats in certain fjords around Maniitsoq (Shearer, 1992). Rapid expansion along the coast followed and from the mid-1960s Faroese and Norwegian fishermen introduced offshore driftnets, followed soon by fishermen from Greenland and Denmark. At around the same time improvements in gear (the introduction of light monofilament nets) enabled fishing in daylight and improved the efficiency of the gear. As a consequence, catches rose quickly reaching a peak of almost 2700 tonnes in 1971. Fishing by non-Greenlandic vessels was phased out in 1972-1975. However, the total catch remained at around 2000 tonnes until 1976 when a TAC of 1190 tonnes was set; the fishery has been regulated since this time. Small catches of salmon are also made on the east coast of Greenland although these are sporadic and restricted by the small number of communities in this area and by drifting polar ice.

Regulatory measures have been agreed by the West Greenland Commission for most of the years since NASCO's establishment (Table 1.2.3.1). These have resulted in greatly reduced allowable catches in the West Greenland fishery, reflecting declining abundance of the contributing salmon stocks. In all but two years since 1998, the fishery has been restricted to an internal-use fishery and commercial export of salmon is not permitted. Catches in the Greenland fishery are presented in Figure 1.2.3.1.


Figure 1.2.3.1. Nominal catch of salmon (tonnes, round fresh weight) in the Greenland salmon fishery, 1960-2018.

The Greenland salmon fishery operates in summer, with a fairly large proportion of the catch commonly being taken in the weeks after the opening of the season in August. Both drift and fixed gillnets continue to operate. The salmon caught in the fishery to the west of Greenland originate in both North America and the Northeast Atlantic. Data on continent of origm in the catch indicate a reasonably even split between fish from North America and Europe in the early 1990s (ICES, 2013). However, the proportion of North American fish in the catch has increased steadily since this time with North American fish comprising 80-90\% of the fish caught in recent years.

The salmon caught at West Greenland are almost exclusively fish in their second stummer at sea, however, these are non-maturing 1SW salmon destined to return to homewaters as 2SW, or older, fish. Fish from all parts of North America are taken in the fishery, while it is primarily only potential MSW salmon from southern countries in Europe (UK, Ireland and France) that are exploited here. Very few salmon of farmed origin appear in the catches at Greenland, and these are not taken into account in assessments.

Table 1.2.3.1. Summary of Regulatory Measures agreed by NASCO for the West Greenland Salmon Fishery (courtesy of NASCO).

| Year | Allowable catch (tonnes) | Comments/other details in the measures |
| :---: | :---: | :---: |
| 1984 | 870 |  |
| 1985 | - | Greenlandic authorities unilaterally established quota of 852 t . |
| 1986 | 850 | Catch limit adjusted for season commencing after 1 August. |
| 1987 | 850 | Catch limit adjusted for season commencing after 1 August. |
| 1988-1990 | 2520 | Annual catch in any year not to exceed annual average (840t) by more than $10 \%$. Catch limit adjusted for season commencing after 1 August. |
| 1991 | - | $\text { Greenlandic authorities unilaterally established quota of } 840 \mathrm{t} \text {. }$ |
| 1992 | - | No TAC imposed by Greenlandic authorities but if the catch in first 14 days of the season had been higher compared to the previous year a TAC would have been imposed. |
| 1993 | 213 | An agreement detailing a mechanism for establishing annual quota in each of the years 1993 to 1997 was adopted by the Comnission. |
| 1994 | 159 |  |
| 1995 | 77 |  |
| 1996 |  | Greenlandic authorities unilaterally established a quota of 174 |
| 1997 |  | An addendum to the 1993 Agreement was agreed by the Commission. |
|  | Internal | Amount for internal consumption in Greenland has been estimated in the past to be 20 t . |
|  | consumption fishery only | Amount for internal consumption in Greenland has been estimated in the past to be 20 t . |
|  | Internal consumption fishery only | Amount for internal consumption in Greenland has been estimated in the past to be 20 t . <br> A Resolution Regarding the Fishing of Salmon at West Greenland was agreed by the Commission. |
| 2001 | 28-200 | Under an ad hoc management programme the allowable catch will be determined on the basis of CPUE data obtained during the fishery. |
| 2002 | 20-55 | Under an ad hoc management programme the allowable catch will be determined on the basis of CPUE data obtained during the fishery. |
| 2003-2008 | Internal | Amount for internal consumption in Greenland has been |


| Year | Allowable catch (tonnes) | Comments/other details in the measures |
| :---: | :---: | :---: |
|  | consumption fishery only | estimated in the past to be 20 t . |
| $\begin{array}{ll} \hline 2009 & - \\ 2011 & \end{array}$ | Internal consumption fishery only | Amount for internal consumption in Greenland has been estimated in the past to be 20 t . |
| 2012-2014 | Internal consumption fishery only | Amount for internal consumption in Greenland has been estimated in the past to be 20 t . |
| $\begin{array}{ll} \hline 2015 & - \\ 2017 & \end{array}$ | Internal consumption fishery only. Greenland unilaterally committed to limit the total annual catch for all components of the fishery to 45 t in 2015, 2016 and 2017. | The fishery will open no earlier than 1 August and close no later than 31 October each year. Any overharvest in one year will result in an equal reduction in the catch limit the following year. Efforts will be made to identify and implement temporal or spatial harvest restrictions that would provide increased protection for weaker stocks Greenland will further improve the monitoring, management control and surveillance of its salmon fishery in accordance with the Plan for Implementation and Control Measures in the Salmon Fishery at West Greenland with the objective of achieving full catch accountability. All Members of the Commission will implement the six tenets. <br> will inform NASCO, in a timely manner, of any modifications to the management of the West Greenland salmon fishery, of the outcome of the 2015, 2016 and 2017 fisheries and of progress with the implementation and effectiveness of its Plan for Implementation of Monitoring and Control Measures in the Salmon Fishery at West Greenland. <br> States of origin will explore opportunities to share experiences with Greenland on monitoring, management control and surveillance in the salmon fishery. <br> A quota of 32 tonnes was set for 2016 fishery. The Commission agreed to review the measure prior to the 2017 fishery. |
| $\begin{array}{ll} 2018 & - \\ 2020 & \end{array}$ | Internal consumption fishery only. Greenland committed to limit the total annual catch for all components of the fishery to 30 t in 2018, 2019 and 2020. | Greenland agreed to prohibit the export of wild salmon or salmon products from Greenland and to prohibit landings and sales to fish processing factories. <br> The fishery will open no earlier than 15 August and close no later than 31 October each year. <br> An annual quota of 30 t was set for all components of the fishery. Any overharvest in one year will result in an equal reduction in the catch limit the following year with no carry forward for any under-harvest into a future year. <br> Greenland will inform NASCO annually of any modifications to the management fishery and will report on progress with |


| Year | Allowable catch (tonnes) | Comments/other details in the measures |
| :---: | :---: | :---: |
|  |  | the implementation and effectiveness of its Plan for Implementation of Monitoring and Control Measures for the fishery. <br> States of origin agree to share experiences on monitoring, management, control and surveillance in the salmon fishery through knowledge-sharing exchange programmes. <br> Greenland will annually collect and verify catch data for all licensed fishers. <br> All fishers for Atlantic salmon will have a licence to fish. Fishing for Atlantic salmon without a license will be prohibited. Only licensed full-time hunters and fishers will be allowed to sell Atlantic salmon at openair markets. All licensed fishers must provide a full accounting of fishing activity and harvest. Fishers who do not provided a full accounting of their catches, including reports of zero catches, within one month of the end of the fishing season at the latest will be prohibited from acquiring a licence for the following season until the required reporting is received. <br> The regulatory measure will also apply to the 2019 and 2020 fisheries unless any member of the West Greenland Commission of NASCO requests its reconsideration or the Framework of Indicators indicates that there has been a significant change to the indicators and, therefore, a ent is warranted. |

Over the past 20 to 30 years, there has been a marked decline in the abundance of Atlantic salmon across the species' distributional range. Wild Atlantic salmon populations are declining across most of their home range and, in some cases, disappearing (1CES, 2008). Generally, populations on the southern edge of the distribution seem to have suffered the greatest decline (Parrish et al., 1998; Jonsson and Jonsson, 2009; Vøllestad et al., 2009), which may be linked to climatic factors. The decline in salmon abundance has coincided with a variety of environmental changes linked to an increase in greenhouse gases and a corresponding increase in temperatures (IPCC, 2001), which is most likely to have manifest effects at the edge of the species range. However, these areas are often also the ones with higher human population densities and therefore, typically, where potential impacts on the freshwater environment may also be greater. A range of factors in freshwater are known to affect stocks including, for example, contaminants, river obstructions, and changing river flows and temperatures (ICES, 2009b; 2010b; Russell et al., 2012). Such factors have potential implications for the survival of juvenile salmon and their resulting fitness when they migrate to sea as smolts (e.g. Fairchild et al., 2002).

Atlantic salmon occupy three aquatic habitats during their life cycle: freshwater, estuarine and marine. Similar factors contribute to mortality in each of these habitats -
competition, predation and environmental factors - but despite occurring in different habitats, these are not independent. Conditions experienced within the freshwater environment can affect the survival of emigrating smolts and marine conditions may subsequently modify the spawning success of fish in freshwater.

The decline in salmon populations has occurred despite significant reductions in exploitation, although this does not preclude possible fishery effects. An underlying cause has been a marked increase in the natural mortality of salmon at sea; the proportion of fish surviving between the smolts' seaward migration and their return to freshwater as adult fish (e.g. Peyronnet et al., 2008; Chaput, 2012). For many stocks, return rates are now at the lowest levels in the time-series, even after the closure of marine fisheries. This reduced survival is thought to reflect climatic factors and broadscale changes in ocean ecosystems as well as factors in freshwater. The exact processes controlling marine survival are relatively poorly understood (Friedland, 1998), although there is growing support for the hypotheses that survival and recruitment is mediated by growth during the post-smolt year, for European stocks at least (Friedland et al., 2009).

Although their habitats are widely separated geøgraphically, thereis strong coherence in recruitment patterns between North American and European stock complexes (Olmos et al. 2019). Recent research suggests recruitment is correlated with ocean temperature variation associated with the Atlantic Muttidecadal Oscillation (AMO) (Friedland et al., 2013). It further appears that thereare differences in the mechanisms affecting stocks in the Northwest and Northeast Atlantic, with ocean climate variability during the first spring months of post-smolt life most important to the survival of North American stocks, while summer climate variation appears to be more important to adult recruitment variation for European stocks (Friedland et al., 2013). It has been speculated that this may be related to the varying roles of predation pressure and sizerelated mortality on the two continental stock complexes.

In addition to changes in climate and potential issues operating in both freshwater and marine environments, various other factors have been postulated as possibly contributing to the decline in stock abundance, including predation, aquaculture impacts and the effects of fisheries. Huge increases in aquaculture production of Atlantic salmon over recent decades (see Section 2.2.1 of the WGNAS report) have created some concerns for wild populations. The main potential impacts include: (i) genetic impacts on wild fish; (ii) discharge of organic material and other wastes; (iii) transmission of diseases and parasites (particularly sea lice) to wild populations; and (iv) concerns about obtaining adequate feed resources from an already heavily exploited marine ecosystem. For example, recent investigations in Norway have demonstrated that gene pools of wild salmon populations in a number of rivers have been gradually changed through introgression of genetic material from escaped farmed salmon (Glover et al., 2012; Glover et al., 2013). Sea lice also continue to be regarded as a serious problem for wild salmonids (Skilbrei et al., 2013; Krkošek et al., 2013) affecting their survival and perhaps also their life-history characteristics (Vollset et al., 2014).

As well as declines in abundance, changes in salmon life histories are also widely reported throughout the geographic range of the species, affecting factors such as sea
age composition, size at age, age-at-maturity, condition, sex ratio and growth rates (e.g. Nicieza and Braña, 1993; Hutchings and Jones, 1998; Niemelä et al., 2006; Peyronnet et al., 2007; Aprahamian et al., 2008; Todd et al., 2008). Changes are also manifest in freshwater stages, affecting factors such as the size and growth of parr and the age of smolting (e.g. Davidson and Hazelwood, 2005; Jutila et al., 2006) and run timing (Kennedy and Crozier, 2010; Otero et al., 2013).

Data

### 2.1 Introduction

Assessment of Atlantic salmon differs from the approaches commonly adopted for other species, for example in respect of the need for at sea surveys and collection of commercial catch per unit of effort (CPUE) data. Instead, the assessment of salmon is based mainly on data collected on individual river stocks (e.g. catches and counts of returning fish), which are raised and aggregated to provide estimates of the number of fish returning to homewaters for different stock groupings. These estimates are used, in turn, to estimate abundance at earlier points in the life cycle of the fish and to inform the development of catch advice.

The provision of management advice for the mixed-stock fisheries at Faroes and West Greenland is based on assessments of the status of stocks at broad geographic scales. The North American Commission (NAC) area is divided into six management units, and the Northeast Atlantic (NEAC) Commission area is divided into 19 regions. Assessment of the status of the stocks in these areas is based on estimates of the total abundance - the pre-fishery abundance (PFA) - of different cohorts of salmon at a stage before the distant water fisheries operate. PFA is defined as the cohorts of salmon maturing as 1SW and MSW fish that are alive prior to all the marine fisheries for 1SW salmon (Rago et al, 1993a). The catch advice for the NEAC area is then provided for the northern (N-NEAC) and southern (S-NEAC) stock complexes and for countries.

The models to estimate the PFA of salmon from different areas are typically based on the catch in numbers of one-sea-winter (1SW) and multi-sea-winter (MSW) salmon in each country or region, which are then raised to take account of estimates of nonreported catches and exploitation rates on the two age groups. In some cases, particularly in the NAC area, returns to homewater are estimated by alternative methods, such as counts at fishways and counting fences, or from mark and recapture dies. The estimates of fish numbers returning to homewaters are then raised to take account of the natural mortality $(\mathrm{M})$ between the date that the fish are deemed to recruit to the particular fishery of interest and the midpoint of the timing of the respective national fisheries. A value of 0.03 per month is assumed for M (Section 3.2.3). The date of recruitment of NAC stocks (and thus the PFA date) is taken as 1 August in the second summer at sea because these fish are first exploited in the distant water fishery at West Greenland. However, NEAC stocks recruit to the Faroes fishery during their first sea winter and so PFA is calculated at 1 January (i.e. seven months earlier) for these stocks.

### 2.2 Input data for assessments-NEAC area

PFA for NEAC stocks is estimated using the run-reconstruction approach described by Potter et al. (2004). The model estimates the PFA of both maturing and non-maturing

1SW salmon because both stock components may be caught in the Faroes fishery (when operating), and data for both the Faroes and West Greenland fisheries are incorporated into the model.

In order to run the NEAC PFA model, most countries provide time-series (beginning in 1971) of catch in numbers, non-reporting rates and exploitation rates for 1SW and MSW salmon. Best estimates and a measure of the uncertainty or error are provided for the non-reporting and exploitation rate data in order to obtain a measure of the uncertainty in the PFA estimates, since these data are commonly derived from expert opinion. In UK (N. Ireland), the PFA model now uses estimates of the numbers of returning adults, split by sea-age. These data are derived from monitored rivers. The latest data input variables used in running the NEAC assessment are listed at Appendix 3.

In some instances, the above information has been supplied in two or more regional blocks per country. In these instances, the model outputs are provided for these regional blocks and also combined to provide one set of outputdata for the country as a whole.

The input data for Finland consists solely of catches from the River Tana/Teno. These comprise both Finnish and Norwegian net and rod catches, as the river marks the border between these countries. The Norwegian catches from the river are not included in the input data for Norway.

Where possible, when the input data are themselves derived from other data sources, the raw data are included in the model. This allows the uncertainty in these analyses to be incorporated into the modelling apprøach. Thus, the catch and sample data used to estimate the catches of Scottish fish in the northeast English coastal fishery are incorporated into the assessments for both UK (England and Wales) and UK (Scotland). For Greenland, catch data are input in the form of harvests (reported and unreported) in weight, along with data from the West Greenland sampling programme.

Descriptions of how the model input data have been derived are presented below for different countries (updated from Crozier et al., 2003; ICES, 2002). The methods used to derive the RFA input data for NEAC countries and options for improving the data also diseussed in Crozier et al. (2003).

## Median dates of return to homewater fisheries

NEAC stocks recruit to the Faroes fishery during their first sea winter and so the date of recruitment (and thus the PFA date) is calculated at January 1st. In deriving PFA from the estimates of fish numbers returning to homewaters, it is necessary to take account of natural mortality between the date that the fish recruit to the particular fishery of interest and the midpoint of the timing of the respective national fisheries. The median return date for 1SW and MSW fish for each country/region are provided in the table below. Thus there is about a six to nine month period between the PFA date and the median time of return to homewaters for maturing 1SW fish and 17 to 20 months for non-maturing fish.

Table 2.2.1.1. Midpoint of recruitment to homewater fisheries for NEAC countries/regions.


Catch: The catch input to the model of Finland represents an estimate based on catch enquiries and the total number of licences issued. The Norwegian catch from the River Teno has been included in the Finnish catch, which results in a set of input data that effectively represents a single river system. Catch composition is estimated based on catch samples and corresponding scale analyses.

Level of unreported catch: Unreported catch is estimated by extrapolating the catches of the fishermen that failed to report their catches, as reporting is not mandatory.

Exploitation rates: Exploitation rates in the river fisheries are derived from radio tagging studies in 1992-1993 and 1995, when 70-100 adult fish (1SW and MSW) were tagged yearly in the estuary. Most of the important river fisheries were covered by these experiments.

### 2.2.2.2 Norway

Area split: Salmon catches in Norway are split into four regions on the basis of climatic and oceanographic differences among the areas. The regions are: (1) southeast Norway from the Swedish border to the border between Rogaland and Hordaland counties, (2) southwest Norway from the border between Rogaland and Hordaland counties to Stad (3) mid Norway from Stad to Lofoten, and (4) north Norway from Lofoten to the border with Russia.

Catch: Nominal catches of salmon in the four regions are used. In recent years there have been improvements in declaring catches. From 1979 there was a weight split 1 SW/MSW ( $<3 \mathrm{~kg} />3 \mathrm{~kg}$ ). From 1993 the split was changed to $1 \mathrm{SW} / 2 \mathrm{SW} / 3 \mathrm{SW}(<3 \mathrm{~kg} / 3-$ $7 \mathrm{~kg} />7 \mathrm{~kg}$ ). Mean weight was provided for most groups and used to estimate numbers in the early part of the time-series, but in recent years the reported nominal catch (reported number of killed salmon in river and sea fisheries summed) is being used. In the input to the PFA model, salmon smaller than 3 kg are regarded as 1 SW fish, whereas salmon larger than 3 kg are regarded as MSW fish. The two largest size groups are thus summed into MSW salmon. In the PFA model input the Norwegian catch data for the River Teno have been removed from the Norwegian catches and incorporated in the Finnish catches.

Unreported catch: No systematic method is used to estimate unreported catches. Inputs are guesstimates based on occasional reports from test fishing, surveillance reports, and questionnaires. There is no evidence that the level of anreported catches differs between the four regions. These estimates are provided by the management authorities.

Exploitation rates: The rates for the national model are guesstimates. For parts of southeast and southwest Norway they are derived from estimated marine exploitation rates from the River Imsa and the River Drammen, respectively. In recent years (from 2009 onwards) exploitation rates for many rivers ( $>50$ ) have been taken into consideration. These exploitation rates haye been obtained using a multitude of methods, mainly from drift counts of spawners or results from counting facilities combined with reported catches in the rivers. The exploitation rates have been adjusted in relation to reduced fishing effort. At present different exploitation levels are used for the different regions, reflecting different harvest regimes in the regions.

For Norway, only data from 1983 onwards have been used for assessment purposes.

### 2.2.2.3 Russia

Area split: The Atlantic salmon rivers of northwest Russia are split into the following four regions: Kola Peninsula - Barents Sea basin; Kola Peninsula - White Sea basin; Archangelsk Region and the Karelia; and the Pechora River region. The split is based on four regions with separate catch statistics and different biological characteristics of the stocks. For example, the difference in age composition and relative abundance of summer and autumn salmon evident among these four regions has influenced the split.

Catch: The declared catch data, in numbers, is available for the full time period (1971 onwards) for all four regions. Catches were allocated to 1SW or MSW age groups on the basis of commercial and scientific catch sampling programmes.

Level of unreported catch: Unreported catches in legal fisheries are estimated from logbooks and catch statistics, by comparing catch survey results with reported catch. Illegal catch is guesstimated and based on local knowledge of fisheries. The major component of the illegal catch in the Barents Sea basin (Kola Peninsula and Pechora River) comes from in-river fisheries and a considerable part of the illegal catch in the White Sea basin (Kola Peninsula and Archangelsk region) comes from coastal areas and this contributes the greatest uncertainties. There is a particular problem with illegal catches on the Pechora River where scientific sampling programmes suggest
that the illegal catch on this river is very high. The level of non-reporting increased considerably in the early 1990s due to the economic changes in Russia and temporary reduction of control and enforcement. Since the late 2000s the higher level of nonreporting occurred in recreational fisheries due to unclear legislation for reporting. All these factors have been considered in deriving the level of unreported catch for the PFA model.

Exploitation rates: Information on exploitation rates is derived from several fisheries in the Kola Peninsula where counting fences are operated and from mark-recapture exercises on the rivers with recreational fisheries. Exploitation rates in Archangelsk and Pechora are guesstimated. These are the basis of the inputs to the model, regional sea age differences being adjusted on the basis of local knowledge from estimated stock levels.

### 2.2.2.4 Sweden

Catch: The catch input to the model is based on annual reported commercial salmon catch on the Swedish west coast, and on voluntary reporting from sport fishing in rivers. This reporting is detailed and considered accurate and is handled by the government agency "Swedish Agency for Marine and Water Management" (commercial catches) and the Swedish University of Agricultural Sciences (noncommercial catches). Unfortunately, reporting of catches from non-commercial fishing for salmon with gillnets or rod and line on the coast is lacking. However, due to fishing regulations these catches are small (permits required for trapnets, ban on gillnets in deeper waters, restrictions on the use of gillnets in shallow waters, limited fishing season, large marine protected areas, ban on selling fish, etc.).

There is a large proportion (mean $64 \%$ for 2002-2018) of reared fish in catches and stocks as a result of compensatory releases of reared smolts (ranching). As all ranched salmon are finclipped the catches of reared fish can be treated separately in the catch statistics. In the reporting from the commercial fishing the catch is not separated into wild and reared fish. The proportion of wild salmon is instead estimated from catch statistics in nearby rivers. Stocking of reared salmon is done in three rivers; two of these wild stocks in tributaries.

Catch and release is practised in most rivers (only rod and line fishing allowed in rivers) but the extent of $C \& R$ is not always known. In most rivers a proportion of the fish is released back alive but any subsequent mortality is not accounted for.

Level of unreported catch: Unreported catch, i.e. non-commercial catch of salmon in the coastal area with gillnets and rod and line, is estimated from guesstimates based on expert judgement from regional fishery officers and the Swedish University of Agricultural Sciences. These estimates are supported with catch inventories carried out in 1999 (Thörnqvist, unpublished), 2004 (Swedish Agency for Marine and Water Management), 2008 (Thörnqvist, unpubl.). Generally, the unreported catch is estimated to be $5-10 \%$ of the reported catch.

Exploitation rates: Few fish counters are present and tagging data exist mainly for reared stocks, where the fishing pressure is higher than for wild stocks. Input for the PFA model is based on guesstimates. In the index River Ätran, data on size and composition of the spawning run and estimates of exploitation are being developed.

Since 2000, a fish ladder with an automatic counter has provided data on the spawning run in this river. Counter data in combination with results from small-scale tagging in this river are used to provide estimates of exploitation rates. One problem is that exploitation rates differ considerably between rivers. During the period 2000-2014 the average exploitation rates for the Swedish stock as a whole have been estimated at $34 \%$ for 1 SW and $39 \%$ for MSW. The exploitation rate increased in 2011-2014 due to increased gillnet fishing on the coast. This fishery has since been closed and exploitation is expected to decrease in future.

### 2.2.2.5 Iceland

Area split: The input data for the PFA model is divided into two a areas. Rivers in the west and south of Iceland are combined into one area and rivers in the north and east into another. This is done on the basis of historic tag recoyeries in ocean fisheries (which occurred in different areas) and different climate and oceanic conditions affecting the salmon life cycle, e.g. run-timing, smolt age, and sea age. Thesouthern and western parts of Iceland fall within the NEAC southern area, while the northern and eastern parts of Iceland fall within the NEAC northern are

Catch: Age-class information is available from individual catch records from logbooks used in the rod fishery. The division into sea age classes is based on a bimodal weight distribution. The 1 SW females are $<3.5 \mathrm{~kg}$ and $2 S W$ females $>3.5 \mathrm{~kg}$, while 1 SW males are $<4 \mathrm{~kg}$ and $2 \mathrm{SW}>4 \mathrm{~kg}$. Scale analyses have shown that the presence of salmon having spent more than two winters at sea and of previous spawners is uncommon and that the categorisation into 1 SW and 2 SW age classes by weight is as fairly accurate. The net catches are recorded on a daily basis. The age split in the net fishery is derived from the weight distribution in the rod fishery from the same river system or from rivers in the same area.

In the River Ranga in southern Iceland substantial smolt releases have occurred since the early 1990s and have now reached a level of 300000 to 500000 smolts annually. Originally, the River Ranga had a small salmon stock with an annual catch of 10 to 90 fish until 1990. The riyer has very limited habitat for salmon production, but these 'ranched' fish now support a substantial rod fishery. The catch in the River Ranga comprised $23 \%$ (18-27\%) of the total reported salmon rod catch in Iceland between 2009 and 2013. Since these fish are expected to have very low spawning success in the river they are excluded from the PFA catch input data.

Level of unreported catch: The fishing rights in Icelandic salmon rivers belong to landowners who must, by law, form a fishery association to manage the fishing right. The rod fishing rights are leased to the highest bidder. No ocean or estuary fisheries are allowed. The unreported catch was originally believed to be low with a guesstimate value of $2 \%$ applied. With increased use of midwater trawls in pelagic fisheries off the coast of Iceland, new information was provided which suggested an increased level of salmon bycatch. Based on a questionnaire survey, the value of unreported catch was therefore revised after 1995 to a value of $10 \%$ of the declared salmon catch. However, more recent analyses of DNA, as well as scale analyses, from salmon sampled as bycatch by Icelandic fishing vessels, indicates a low percentage of Icelandic salmon. Based on this, and other available information, a new estimate of unreported catch is
now applied for Iceland at 4\% of the declared catch for 1SW and MSW salmon since 1995.

Exploitation rates: Rates of rod exploitation are based on rivers with fish counters and catch records from logbooks. The estimates of exploitation are 40-50\% for 1SW salmon and $50 \%$ to over $70 \%$ for 2 SW salmon. The exploitation estimate for an in-river gillnet fishery is $39 \%$ to $52 \%$, with a higher exploitation rate on larger fish. Information on the number of fish subject to catch and release in rod fisheries is also available from logbooks. The proportion of released fish has been increasing since 1996. The reduced exploitation due to catch and release is taken into account in the annual estimate of exploitation for both 1 SW and 2 SW stock components in the PFA model inputs.

Median return date of 1SW and MSW: Run timing can vary both between years and between areas. The median return date of 1SW and 2SW salmon in south and west Iceland is mid-June and early June respectively. The median date of return is later in the north and east of Iceland, mid-June for MSW and early July for 1SW salmon.

### 2.2.2.6 Denmark

The Working Group collects and routinely reports the annual catch of salmon taken in Denmark. However, the small Danish catches are not included in the assessment process used in developing catch advice for the distant water fisheries.

Catch: The catch input is based on continưously collected reports of salmon taken in the recreational fishery in Danish west coast streams (from Internet sources), which all hold populations of wild salmon. In four of these, where salmon populations have always been found, there is a large proportion of reared (finclipped) salmon in the catch, but these are all F1 (second generation) offspring from the original populations. In the one catchment in eastern Denmark (Gudenå), where the salmon population is not genetically native to the stream, the annual catch is guesstimated.

Level of unreported catch: Unreported catch is expected to be negligible in the western streams because the fishing is closely regulated and controlled by the anglers. In the eastern stream (Gudenå) unreported catch is guesstimated.

Exploitation rates. Exploitation rates may be derived from the total catch related to estimates of the total run (calculated by mark-recapture surveys on a three-year cycle
the fourstreams with original populations on the west coast).

## Data inputs for Southern NEAC countries

### 2.2.3.1 France

Catch: The estimation of salmon catch in France comes from two main sources: (1) mandatory declaration of rod and line catches and from the Adour nets operating in the lower river (scales are sampled from each fish caught) to the Centre national d'interprétation des captures de salmonidés migrateurs (CNICS) under Agence Française pour la Biodiversité (AFB); and (2) mandatory declaration of catches made by professional net fishermen to Affaires Maritimes, under the Ministère de la Mer, who since 2008 have delegated responsibility for collection and first processing of catch data to the Regional Boards for Sea Fisheries and Aquaculture Catch. At the same time, catches at sea are declared to the Institut Français de Recherches pour l'Exploitation de
la Mer (Ifremer), who are responsible for archiving and scientific processing of all fisheries data. Salmon catches have not been reliably collated and made available until recently. Since 1985, the 1SW/MSW split has been based on scale interpretation of the in-river catch (based on scale reading) and on a categorisation based on length thresholds for catches in estuaries and at sea. The figures prior to 1985 are not considered as reliable as the later ones.

Level of unreported catch: Unreported legal catch for the rod and line fishery has been estimated by catch inquiries made by environmental inspectors of AFB on each river. These procedures are still operating in some areas, but estimates are considered less reliable in recent years. The estimation of the professional net fishery catch (Adour Basin) is thought to be reliable and no unreported legal catch is considered to apply.

For most years, the unreported illegal catch is not assessed and a minimal value is provided on a precautionary basis. This unreported illegal catch has been assessed in some years by ad hoc inquiries in the estuary of a number of rivers in Brittany (e.g. in 2001) and on the coast (e.g. Baie of Mont Saint-Michel in 2000). The "unreported catch" is included in the nominal catch. No estimates of unreported catch are available for the early part of the time-series (prior to 2001). Thus, the rates input to the model for 1SW and MSW for the early period are near zero and range from 0 to 0.00001 . Higher values in the range $20 \%$ to $40 \%$ for 1 SW and $15 \%$ to $30 \%$ for MSW fish are applied more recently.

Exploitation rates: Exploitation rates are derived from the index River Scorff in Brittany. This is an in-river rate, by rods only, where there are no, or very few, fish thought to be caught on the estuary or coast. Rates are also derived for the Adour river system, where a rough estimation is proyided by using the lower values of adult run estimates through facilities in the three rivers flowing to Adour, and the declared catches on the coast, estuary and river, respectively by nets and rods. Some caution is necessary regarding these rates from the Adour given the uncertainties in the different estimates. The rod catch on the index River Nivelle is very small and the probable net exploitation in the estuary and coast is unknown, so exploitation rates are not used for this system. Some data on exploitation rates are also collected by AFB on the index River Bresle, but sea trout are the dominant angled species in this river.

## 2,2,3.2 Ireland

Catch: The data are derived from annual declared catches within fisheries districts, management units implemented by Regional Fisheries Boards. Since 2007 river and estuarine specific angling and commercial catch data have been complied. The Fisheries Boards were amalgamated into a single body, Inland Fisheries Ireland, in 2010 which currently takes responsibility for compiling catch statistics. Catches are split by age on the basis of a reported age distribution from 1980 to 1988. In the absence of any other information the mean proportion of 2 SW salmon in the series $(7.5 \%)$ has been used since 1988 and a mean of $10 \%$ has been used prior to 1980. Since the introduction of a carcass tagging and logbook scheme for angling and commercial fisheries in Ireland in 2002, sea age classes in the time-series since 2007 have been determined based upon catch dates and weights in accordance with national river stock assessments. The catch is not corrected for returns from releases of smolts for ranching or enhancement but these are not a major component of the catch.

Level of unreported catch: The values are guesstimated from local reports and knowledge achieved during catch sampling and fisheries protection activities.

Exploitation rates: A coded-wire tagging (CWT) programme has been operated in several rivers in Ireland since 1980. Up to 300000 hatchery smolts and up to 5000 wild smolts are tagged and released annually. There is also a substantial dataset on wild salmon from the monitored River Burrishoole, providing a further index of wild returns and exploitation rates. Overall, there are estimates of exploitation rates available for three wild stocks and seven hatchery stocks for both 1SW and 2SW salmon. Up to the closure of the marine mixed-stock fishery in 2006, the annual mean of the 1SW wild exploitation index is used as the input data for the lower range of exploitation in the PFA model while the mean of the 1SW hatchery index is used as the upper range. The annual mean of the 2 SW wild and hatchery exploitation index was used as the input data for the upper and lower range of exploitation in the PFA model depending on which is higher or lower in that year. Since 2006 the main exploitation input has been from the rod catch which is estimated from CWT estimates for some rivers and also rivers with counters.

### 2.2.3.3 UK (England \& Wales)

Catch: Nominal catches for UK (England \& Wales) have been derived from the catch returns submitted by netsmen and anglers and split into 1SW and MSW categories using two different methods. Since 1992, monthly age-weight keys derived from salmon caught in the River Dee trap (an index river) have been used to estimate the age of all rod-caught fish where a weight and date of capture have been provided. This has then been scaled up to the total catch (rods and nets combined) on a pro-rata basis. In earlier years (1971-1991), the age composition of the total catch has been estimated using the mean weight of the fish caught and the mean weight of 1SW and MSW salmon recovered in tagging programmes.

As the contribution of farmed and ranched salmon to the national UK (England \& Wales) Catch is negligible, the occurrence of such fish is ignored in the assessments of the status of national stocks. However, a large proportion of the fish taken in the northeast coast fishery are destined for Scottish rivers, and these are deducted from the returning stock estimate for UK (England \& Wales) and added to the data for UK (Scotlond) in the ICES assessment. This proportion is estimated to have declined from $5 \%$ of the northeast net catch in the early part of the time-series to $75 \%$ in the late 1990s do around $65 \%$ since 2003. This reflects both the steady improvement in the status of the stocks in northeast England and the phase out of the English driftnet fishery since 2003.

Level of unreported catch: All licence holders are required to provide the Environment Agency with details of their catch of salmon and the number of days fished on each river or, for nets, each fishery at the end of the season. Catch returns are received from all net licence holders and from a high proportion of full season anglers, and the latter account for the majority of fish caught in a catchment, typically $96-98 \%$. The main correction for underreporting is therefore currently made in respect of perceived inaccuracies in the returns, although more substantial corrections have applied in the past.

There are few independent measures of underreporting in the rod fishery, but these indicate that the level is currently small. A value of $10 \%$ is applied for correction purposes based on the method of Small (1991). Historically, underreporting was a much more serious problem. As a result of changes in the licensing and associated catch return system covering UK (England \& Wales) in the early 1990s, the percentage of underreporting in the rod catch was estimated to have decreased from $\sim 50 \%$ to $\sim 20 \%$. Since the mid-1990s, awareness campaigns and enhanced catch reminder systems have further reduced underreporting to the levels currently estimated. Since 2015, a new online reporting system and the introduction of 365-day licences (valid from time of purchase) have created some reporting difficulties and larger corrections have been required to account for under-reporting since this time. It is anticipated that this will cease to be necessary with growing familiarity with the new system.

For the net fishery, a figure of $8 \%$ was used from the late 1990 s to 2008 to adjust for the level of underreporting, based on the outcome of surveillance operations. The level may have been substantially higher in the past in certain fisheries, possibly as much as $50 \%$. However, following the successful introduction oflogbooks and a carcass tagging scheme in 2009, there is now considered to be minimal underreporting in net fisheries. A figure of $2 \%$ has been assumed since 2009.

An earlier questionnaire survey of Environment Agency enforcement staff suggested illegal catches were around $12 \%$ of the dectared net and rod catch. However, since the introduction of a carcass tagging scheme and a ban on the sale of rod caught fish in 2009, it has been substantially more difficalt to dispose of illegally caught fish. Since this time, illegal catches have been estimated to have been reduced to $6 \%$ of the declared catch.

Exploitation rates: Exploitation rates for a number of monitored fisheries in UK (England \& Wales) are derived annually. National exploitation rates have then been estimated by deriving time-series of 'standard fishing units' employed in the salmon fisheries for the period 1971 to the present. For the period 1971 to 1997, these are calculated from the numbers of licences issued weighted by their relative catching power, which is estimated from historic CPUE data; and for the period 1998 to the present, they are calculated from the numbers of days fished by different net categories weighted in the same way. The annual exploitation rates are then estimated by referencing the number of 'standard fishing units' employed over the two periods relative to average age-specific exploitation estimates derived for the 1997 and 1998 Additional information: Further details on the derivation of estimates within UK (England \& Wales) are available in the annual stock status reports (e.g. Cefas, Environment Agency and Natural Resources Wales, 2017), available at:
https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attach ment data/file/709614/SalmonAssessmentReport-2017-final.pdf

### 2.2.3.4 UK (Northern Ireland)

Area split: Originally, a single assessment was carried out for UK (Northern Ireland). However, the data used were derived from two fishery management areas (Loughs Agency and DAERA areas), which publish separate catch statistics and have differing fishing regulations. On the basis that stock status in the two areas may differ (Crozier et al., 2003) the two areas were assessed separately from 2001.

Catch: As no commercial fishing has been conducted in the Loughs Agency area since 2010 nor in the DAERA area since 2012 the Northern Ireland catch statistics currently (since 2014) rely solely on rod catches. Overall UK (Northern Ireland) rod catch estimates are available since the introduction of a carcass-tagging scheme in 2001. These catch statistics are used as an input in the model. Estimates of sea age composition of the catch for the time-series are based on 1SW/MSW data from adults returning to the River Bush (an index river).

Level of unreported catch: Estimates of unreported catch, as a result of illegal fishing, are based on intelligence reports from DAERA and Loughs Agency fishery officers. These are guesstimates only, with no verification possible. Annual adjustments in unreported catches have been used since tagging programmesstarted in the mid1980s. Prior to that, a constant under reporting figure is used, as no annual data are available. The introduction of the carcass tagging schemein 2001 has led to a reduction in unreported catches.

Exploitation rates: Estimates of exploitation rates were historically based on the River Bush microtagging programme. Exploitation from this monitored river (which is in the DAERA fishery area) was used as an input figure for all UK (Northern Ireland) fisheries (Loughs Agency and DAERA areas). However, as currently no commercial fishery for salmon exists in the DAERA and Loughs Agency areas, exploitation rates are based on rod exploitation in the DAERA and Loughs Agency alone.

Adult counts: In the DAERA area, counts of adult returns to the rivers Bush and Bann are used as input data for the run-reconstruction model from 2000 onwards. These values are scaled up to the estimated total run using a factor of $1 /(0.67 \pm 0.05)$ for 1 SW 05) for MSW fish, based on more detailed run estimates for 2015. Adult counts are used because these data were more informative about the number of returning adults than very low nominal catches.

Possible improvements: A possible improvement would be to have better data avaitable on sea age composition of all Northern Irish fish. The River Bush and Bann estimates are based on annual scale analysis of a subsample of the adult returners. Since the closure of the Foyle commercial fishery no sea age data are available on Loughs Agency area returning adults and thus this is currently based on estimates from historical data.

In addition, a higher return rate for the carcass tagging scheme would result in more reliable estimates of exploitation rates. Recently the carcass tagging return rate in UK (Northern Ireland) has varied between $14 \%$ and $55 \%$.

### 2.2.3.5 UK (Scotland)

Area split: The country is divided into eleven statistical regions for the purposes of collating and publishing salmon fishery statistics (Marine Scotland Science, 2012). Within the PFA run-reconstruction model, UK (Scotland) is divided into two broad areas (east and west), the split being influenced by the contrasts in topography, river size and the potential migratory routes of post-smolts. The east grouping comprises the East, Northeast, Moray Firth, and North statistical regions, the remaining statistical regions comprise the West grouping in the run-reconstruction model.

Catch: Annual declared catches are collated according to the area split defined above. Reported retained and released catches of wild salmon, taken by net and rod fisheries, are provided separately for two age classes, one sea-winter and multi sea-winter fish. Catch sampling programmes have shown that there is a variable (by region, year, and fishery) proportion of 1SW salmon mis-categorised as MSW salmon in the reported rod catches. The methods used to align abundance estimates at ICES with those used in domestic assessments (see exploitation rates below), also correct for these reporting biases.

Level of unreported catch: Previously the unreported catch ranges used in the national model were based on guesstimates made by local managers in some eastern areas of the country (MAFF, 1991). The differences in the ranges used for the east and west groupings were based on a subjective view that unreported catches in the west area were likely to be greater than in the eastarea due to the geographic differences between the regions. However, in the absence of empirical evidence with which to support these differences, the current unreported catch rate of $10 \%$ was applied throughout the series in both the east and west areas. Error around this value is a subjective estimate of the (normally distributed) uncertainty in the parameter and was set to $5 \%$.
Exploitation rates: Abundance is estimated from total rod catch (retained \& released) raised using a correction factor to align abundance estimates of home water returns from the runreconstruction model with those derived in domestic assessment models. Exploitation rates are estimated as all methods retained catch expressed as a proportion of home water returns for both 1 SW and MSW salmon.

Additional information: Estimates of spawner abundance take into account estimates of catch and release and natural in-river mortality. Little direct evidence of either source of mortality is available for Scotland. Based on limited information from radiotracking studies, the model assumes catch \& release mortality of $10 \%$ (Webb, 1998; Smith, Middlemas, and MacLean, 2014) and an additional in-river mortality of $9 \%$ to count for other factors such as predation and disease (Milner et al., 2000). Fecundity estimates were also revised to align with domestic assessment methods (Marine Scotland Science, 2017). Time series of mean annual eggs per female for both 1SW and MSW salmon were used together with a constant point estimate of sex ratio to estimate egg deposition within the run-reconstruction model. The analysis thus accounts for shifts in fecundity related to observed changes in the lengths of returning fish.

### 2.2.3.6 Spain

The Working Group collects and routinely reports the annual catch of salmon taken in the recreational rod fisheries in Spain (mainly Asturias). However, the small Spanish catches are not included in the assessment process used in developing catch advice for the distant water fisheries.

### 2.2.4 Data inputs for Faroes and West Greenland fisheries

### 2.2.4.1 Faroes

Reported catch: The Faroes fishery has not operated since 2000. When the fishery was being prosecuted, catch data were derived from the landings of salmon caught in the commercial and research fisheries that operated in the Faroes EEZ and the northern Norwegian Sea. Catches for each season (i.e. November in year n to May in year $\mathrm{n}+1$ ) are assigned to the second year (i.e. year $\mathrm{n}+1$ ). These fish are classified into 1SW and MSW age groups according to their age (or potential age) on January 1st during the fishery (i.e. a post-smolt caught in November is classified as 1SW).

Unreported catch: All fish less than 63 cm total length have been discarded in this fishery and so an unreporting rate of $10-15 \%$ (with an error of $+/-5 \%$ ) has been used for 1SW fish; there is thought to have been negligible non-reporting of MSW fish.

Catch composition: Estimates of the proportion of farmed fish in the catch for the period 1981 to 1995 have been derived from scale reading (ICES, 1996, Hansen et al., 1997); prior to 1981 all fish are assumed to have been wild, and since 1997 a value of 0.8 has been used.

Tagged fish originating in North America have been recaptured in the fishery (e.g. ICES, 1991), but excluded from catch advice.

The country of origin of the catch had been estimated based on tagging studies undertaken in the early 1990s (Hansen et al., 1999). These were subsequently replaced by estimates based on genetic analysis.

Genetic investigations, based on salmon scales removed from fish caught in the fishery in the 1980s and 1990s provided an estimated proportion of North American fish in the catches at Faraes. Estimates of $5.7 \%$ (1SW) and 20.5\% (MSW) have been used in subsequent analyses (ICES, 2015).

The composition of the European component was investigated using individual genetic assignments and gave an overall 1SW stock composition of $84.2 \%$ Southern European, 9.0\% Northern Europe, 1.2\% Icelandic and 5.7\% North American (ICES, 2015). The overall composition of the MSW catch was determined as $20.9 \%$ Southern European, $58.0 \%$ Northern Europe, $0.6 \%$ Icelandic and $20.5 \%$ North American (ICES, 2015).

It was not possible to use the genetic assignments to estimate the composition of the catches to country/regional level, but they suggested that the composition within the stock complexes was broadly similar to the relative proportions of the PFA estimates and so the breakdown of catches at this level can be made by applying the relative proportions of PFA (ICES, 2015). Sources of uncertainty in these estimates are described in ICES (2015).

### 2.2.4.2 West Greenland

Catch: The total nominal catch (i.e. tonnes round fresh weight) in the West Greenland fishery is reported and converted to numbers using a mean weight obtained from the sampling programme.

Unreported catch: Estimates of unreported catch were not provided for the period from 1993 to 1999; an annual estimate of non-reported catch, varying from 5 to 20 tonnes was provided by the Greenland representative. Since 2000 a nominal figure of 10 t per year has been provided.

Efforts have been made to provide further information on the level of unreported catch at West Greenland. Since 2002, some assessment of the unreported catch, primarily for commercial landings, has been provided by comparing the weight of salmon seen by samplers involved in the international sampling programme and the corresponding community-specific reported landings. However, since sampling only occurs during a portion of the fishing season, these are considered to be minimum estimates for unreported catch. In addition, there is currently no quantitative approach for estimating the unreported catch for the private fishery. A telephone survey of fishers was carried out following the 2014 to 2016 seasons and provisional findings were provided to WGNAS in 2015 to 2017. These are discussed further in Section 51 of the Working Group report (see above); such investigations may provide a basis for revising estimates of unreported catch in future.

Continent of origin: The catch at West Greenland was divided into NAC and NEAC components using scale characteristics until around 2000 and since that time genetic analysis has been used. For the period when scale characteristics were used, the input data to the model are the minimum and-maximum estimates of the proportion of NAC fish (from which minimum and maximum proportions of NEAC fish are calculated). For the subsequent period, the inputs are the numbers of NAC and NEAC fish identified in the samples.

### 2.2.5 Improvements to NEAC input data

NEAC countries have made ongoing efforts to improve the input data used in assessments Modifications to input variables are reported by WGNAS in the year in which they are first implemented.
Over recent years, efforts have been made to reduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes). However, the methods used to derive estimates of unreported catch vary markedly between countries. For example, some
untries iñclude only illegally caught fish in the unreported catch, while other
countries include estimates of unreported catch by legal gear as well as illegal catches in their estimates.

Descriptions of the national approaches used for evaluating unreported catches have been reported at various WGNAS meetings (e.g. ICES, 1996; 2000; 2002; 2010a). In addition, detailed reports describing national procedures for evaluating illegal and unreported catch, and efforts to minimise this, were submitted by parties to NASCO in 2007 in support of a special theme session on this issue. Full details are available at: http://www.nasco.int/pdf/2007\ papers/CNL(07)26.pdf.

Input data commonly rely on rod catches and the practice of catch and release has become increasingly important in recent years to reduce levels of exploitation on stocks. In the NEAC areas, catch and release estimates from the rod fisheries are not available from all countries and, when they are, corrections for catch and release
mortality are commonly not applied. As the practice of catch and release is increasing, WGNAS have previously recommended (ICES, 2010a) that consideration should be given to incorporating mortality associated with this practice in river-specific, regional and national assessments.

The procedures currently used to incorporate catch and release and unreported catches into regional, national and international assessments are summarised in Appendix 1 (from ICES, 2010a).

One weakness of the NEAC model is that it is heavily dependent upon catch data and estimates of exploitation rate. In most salmon fisheries in the NEAC area, more than half the catch is reported, and in many cases it approaches $100 \%$. However, as stocks have declined, exploitation rates have been reduced to very low levels, and estimates of abundance are therefore becoming increasingly sensitive to this parameter. This inevitably means that uncertainty in the estimates is increasing, and it therefore strengthens the need to make use of alternate sources of information on stock abundance, such as adult counts.

### 2.3 Input data for assessments-NAC area

The run-reconstruction model for NAC developed by Rago et al. (1993a) is used to estimate the PFA of non-maturing 1SW salmon of North American origin (beginning in 1971). Only the West Greenland fishery is of relevance in the context of distant water exploitation of NAC stocks. This fishery exploits predominantly (>95\%) 1SW nonmaturing salmon (destined to return primarily as 2SW salmon) and hence it is only necessary to estimate the andance of this age group prior to the fishery at Greenland. The other fish taken in the fishery represent 2SW and older non-maturing salmon and previous spawners (ICES, 2003) However, PFA estimates for maturing 1SW salmon as well as large salmon (containing all MSW age groups of salmon including repeat spawners) are derived from the run-reconstruction model.

The starting point for the reconstruction requires estimation of the returns of 2 SW salmon, small salmon and large salmon to the six regions in eastern North America: Newfoundland, Québec, Gulf, Scotia-Fundy, and USA. With the progressive closure of commercial fisheries (1984 for the Gulf and Scotia-Fundy regions; 1992 for Newfoundland; and 1998-2000 for Labrador and Québec) abundance estimates have led less. on harvests and increasingly on estimated returns to rivers raised to production areas. The returns for each region are estimated with the uncertainty defined by a range of minimum and maximum values based on the best information available for each region (Chaput et al., 2005).

The annual pre-fishery abundance of non-maturing 1SW fish for year i, destined to be 2SW returns (excluding 3SW and previous spawners), represents the estimated number of salmon at West Greenland prior to the start of the fishery on August 1st. Definitions of the input variables used in the model are given in Table 2.3.1. The PFA estimate is constructed by summing 2SW returns in year $\mathrm{i}+1$ [NR2(i+1)], 2SW salmon catches in commercial and indigenous peoples' food fisheries in Canada [NC2(i+1)], and catches in year i from fisheries on non-maturing 1SW salmon in Canada [NC1(i)] and Greenland [NG1(i)].

Table 2.3.1. Definitions of key variables used in continental run-reconstruction models for North American salmon.

| i | Index for PFA year corresponding to the year of the fishery on 1SW salmon in Greenland and Canada |
| :---: | :---: |
| M | Natural mortality rate (0.03 per month) |
| t1 | Time between the midpoint of the Canadian fishery and return to river $=1$ month |
| S1 | Survival of 1SW salmon between the homewater fishery and return to river \{exp-M * t1\} |
| H_s(i) | Number of "Small" salmon caught in Canada in year i; fish $<2.7 \mathrm{~kg}$ |
| H_l(i) | Number of "Large" salmon caught in Canada in year i, fish $>=2.7 \mathrm{~kg}$ |
| AH_s | Indigenous and resident food harvests of small salmon in northern Labrador |
| AH_1 | Indigenous and resident food harvest of large salmon innorthern Labrador |
| f_imm | Fraction of 1SW salmon that are immature, i.e. non-maturing: range $=0.1$ to 0.2 |
| af_imm | Fraction of 1SW salmon that are immature in indigenous and resident food fisheries in northern Labrador |
| q | Fraction of 1SW salmon present in the large size market category; range $=0.1$ to 0.3 |
| MC1(i) | Harvest of maturing 1SW salmon in Newfoundland and Labrador in year i |
| i+1 | Year of fishery on 2SW salmon in Canada |
| MR1(i) | Return estimates of maturing 1SW salmon in Atlantic Canada in year i |
| NN1(i) | Pre-fishery abundance (PFA) of non-maturing 1SW + maturing 2SW salmon in year |
| $\mathrm{NR}(\mathrm{i}) \quad$ Return estimates of non-maturing + maturing 2SW salmon in year |  |
| NR2(i+1) |  |
| NC1(i) Haryest of non-maturing 1SW salmon in Nf |  |
| NC2(i+1) Harvest of maturing 2SW salmon in Canada |  |
| NG(i) Catch of 1SW North American origin salmon at Greenland |  |
|  | Time between the start of the fishery at West Greenland (August 1) and return othe coast of North America $=10$ months |
|  | Survival of 2 SW salmon between August 1 (at West Greenland) and return to the coast of North America $\left\{\exp -\mathrm{M}^{*}\right.$ t 2$\}$ |
| (1) | Pre-fishery abundance of maturing 1SW salmon in year i |

The latest data input variables used in running the NAC assessment are listed at Appendix 4. More detailed descriptions of how the model input data have been derived for each region of North America are presented below.

### 2.3.1.1 Labrador

For Labrador stocks, it was thought to be inappropriate to develop total recruits from angling catches and exploitation rates similar to techniques used for rivers in insular Newfoundland. The problem with using angling catches to derive returns for Labrador is, that until 1994, there were no estimates of exploitation rates available other than for the salmon population of Sand Hill River and these were 20 years out of date. Also,
because Labrador coastal rivers are isolated, the exploitation rates are low and highly variable depending on the presence of an angling camp and its success in attracting guests as well as the nearness of local communities. Thus, exploitation rates would depend, and vary from one year to the next, on the success of angling camps in attracting anglers and may not be applicable to other Labrador rivers. Thus, all estimates of returns and spawners until 1998 were based on commercial catches as the only source of usable continuous time-series of data.

## Before 1998

The general approach is to use exploitation rates to convert commercial catches of small and large salmon in Labrador to total population prior to the commercial fishery. River returns and spawners were estimated by subtracting the commer ial catch from these populations, and accounting for non-Labrador interceptions. The estimated number of Labrador origin large returns is calculated as:

$$
\mathrm{LR}=(\mathrm{CC} * \mathrm{PL}) / \mathrm{u}
$$

where,
$\mathrm{LR}=$ Labrador returns, $\mathrm{PL}=$ proportion L and $u=$ exploitation rate

The estimated number of Labrador origin small returns is determined from equation (1) but using commercial catches of small salmon.

Parameter values for sea age and the proportion of salmon of Labrador origin comes from the sampling program in the commercial fishery, 1974-1991. In 1997, commercial sampling resumed with samples being collected throughout the fishery at Makkovik and Rigolet in SFA 1 and Cartwright and St Lewis/Fox Harbour in SFA 2.

River age distribution of commercial samples of small and large salmon from Labrador have been found to censist, on average, of about $75-80 \%$ river age 4 and older in SFAs $1 \& 2$. The conmercial samples came from commercial catches sampled in Labrador at several sites along the Labrador coast including Square Islands (SFA 2) and at Nain (SFA 1) (Anon, 1993b). In total, 46320 salmon were sampled for scales and aged. Labrador salmon stocks are thought to contribute about $70 \%$ of the total production of four year, and older, river age salmon, with the other $30 \%$ coming from northern Québec. Thus, when non-Labrador salmon are factored in at $30 \%$ applied to the river age distribution, then $60-80 \%$ of the harvest of small and large salmon (PL) in Labrador are of Labrador origin (Anon, 1993b). In 1997, in SFA 1, the percentage of the commercial catch that was of Labrador origin was for large salmon 68\% (95\% C. I. 64.3$72.5 \%$ ); whereas for small salmon it was $39 \%$ ( $95 \%$ C.I. $35.6 \%-41.6 \%$ ). In 1997, in SFA 2, the percentage of the commercial catch that was Labrador origin was for large salmon $92 \%$ ( $95 \%$ C. I. $88.4-95.2 \%$ ); whereas for small salmon it was $80 \%$ ( $95 \%$ C.I. $74.8 \%-85.0 \%)$.

Exploitation rates (u) were calculated from the smolt tagging study in 1969-1973 on Sand Hill River (Reddin, 1981; Reddin and Dempson, 1989). Exploitation rates of 0.28 to 0.51 for small salmon and 0.83 to 0.97 for large salmon from the tagging study were changed to base exploitation rates of 0.3 to 0.5 on small salmon and 0.7 to 0.9 on large
salmon and were assumed to apply to all of the salmon populations in SFAs 1, 2, and 14B for the period of 1969-1991 (Anon., 1993b). After 1991, due to the Management Plans for the commercial fishery in Labrador and Newfoundland, several changes occurred that would reduce exploitation of Labrador origin salmon. These changes include: (1) reductions in effort as commercial salmon fishermen chose to sell their licences from a buy-out agreement begun in 1992, (2) a moratorium on commercial fishing in Newfoundland that would increase the number of Labrador salmon in Labrador coastal waters, and (3) season reductions due to the varying opening dates and early closures from the quotas applied in 1995 and 1996. The effects of these changes were quantified in the exploitation model as follows:

$$
\mathrm{u}=1-\mathrm{e}^{-\mathrm{aF}}
$$

where: $\mathrm{a}=$ fraction of the 1991 licensed effort remaining in
In 1994-1996, the licensed effort for all of Labrador was $37 \%$ of the 1991 level of 570 licences, in 1993 it was $55 \%$, and in 1992 it was $87 \%$. In any given year, it was assumed that $90 \%$ of licensed fishermen were active. Fishermen reported during public consultations that in 1995 and 1996 many licensed salmon fishermen did not fish for salmon but fished for crab instead. This was verified by Fisheries Officers who reported that of the 218 licensed salmon fishermen only 132 were active in 1996. Another method of obtaining actual effort information is also available since, beginning in 1993 commercial fishing vessel (CFV) numbers have been recorded on sales receipts issued to fishermen by fish plants. Enumeration of licensed salmon fishermen actively fishing was made by determining the number of CFVs in the Statistics Branch catch records. Active effort in 1991 and 1992 was assumed to be $90 \%$ as it was in 1993 and 1994 from the CFV file. Thus, the exploitation rates (u) were modified due to effort reductions in equation (2) using estimated active licences from 1991 as a base and the number of active licences in 1995, 1996 and 1997. The modified exploitation rates (ue) for 19921997 used the licensed effort in equation (2).

The tagging study on Sand Hill River, 1969-1973 showed that Labrador small and large salmon were not only carght in Labrador, but also in the commercial fisheries along the northeast coast of Newfoundland (both small and large) and at West Greenland only) (Anderson, 1985). For small salmon, out of a total of 100 (1SW) tag returns there were 24 from Newfoundland. For large salmon, out of a total of 137 (2SW) tag returns there were 41 from Newfoundland.

1992-1997: the moratorium on commercial fishing in Newfoundland would have released small and large salmon to Labrador. The effect of salmon released from Newfoundland in 1992-1996 was evaluated against the exploitation rates as follows:

$$
\begin{align*}
& \text { un }=\left(1-\left(\left(24^{*}(1-\mathrm{ue})\right) / 100\right)\right)^{*} \text { ue, for small salmon, and } \\
& \text { un }=\left(1-\left(\left(41^{*}(1-\mathrm{ue})\right) / 137\right)\right)^{*} \text { ue, for large salmon } \tag{3}
\end{align*}
$$

The new estimates of fishing mortality (un) in 1992-1994 included adjustments for the closure of the commercial fishery in Newfoundland based on the results of the Sand Hill River tagging study. Season reductions due to the varying opening dates and early closures from the quotas applied in 1995 and 1996. In 1995, adjustments were made to account for the new opening date for the commercial fishery in Labrador of July 3 changed from June 20 the previous year. For 1995, the accumulative effect of these,
weighted to SFA catches, was to reduce the catch so that for small salmon the current catch represents $86.0 \%$ of small salmon and $62.7 \%$ of large salmon. In 1996, the opening date reverted to June 20 but the quota levels resulted in early closures in SFA 2 of 2A July 10, 2B - July 8, and 2C - July 2 while SFA 1 and 14B did not close. For 1996, the accumulative effect of these weighted to SFA catches was to reduce the catch so that for small salmon the current catch represents $53 \%$ of small salmon and $61 \%$ of large salmon. In 1997, the opening date remained at June 20 but the quota levels resulted in early closures in SFA 2 of 2A -July 12, 2B - July 15, and 2C - July 13 while SFA 1 closed on October 15 as the quota was not caught. For 1997, the accumulative effect of these early closures was to reduce the catch so that for small salmon the current catch represents $47 \%$ of small salmon and $64 \%$ of large salmon. The season changes reduce catches and hence lower exploitation rates. The effect of shorter seasons in 1995, 1996 and 1997 was evaluated against the exploitation rates in section B as follows:
$\mathrm{US}=\mathrm{UN} * \mathrm{SC}$, for small salmon, where SC is season change, and $\mathrm{US}=\mathrm{UN} * \mathrm{SC}$, for large salmon

The new estimates of fishing mortality including effort reductions, adjustments for the closure of the commercial fishery in Newfoundland, and shorter seasons due to opening dates and quotas results in the following exploitation rates which were applied to catches. The cumulative effect of factors $A, B$, and $C$ is to reduce exploitation on Labrador origin salmon.

Labrador origin 2SW returns (LR2SW) were derived from eq. 1 by:
LR2SW = LR * P2SW
where: $\mathrm{P} 2 \mathrm{SW}=$ proportion of the large salmon that is 2 SW salmon.
The SR1SW were calculated as in equation (5) but using P1SW which is the proportion of the catch that is 1-sea winter in age and maturing to enter freshwater and spawning in the year of capture. The parameter values for P1SW of 0.1 to 0.2 come from Anon. (1991).

The 2SW component was estimated separately for salmon caught in SFA 1, 2 and 14B. In SFA 1, commercial sampling at Nain of large salmon showed the proportion of 2SW was on average about $84 \%(n=6542), 1977-1991$. Thus, a range of $0.7-0.9$ was used for SFA 1. In SFA 2, commercial sampling of large salmon averaged 69\% (n=4793) 2SW salmon, 1977-1991. There were no commercial samples available for SFA 14B. Thus, for SFAs $2 \&$ 14B a range of $0.6-0.8$ was used. For the 1SW component, commercial samples at Nain in SFA 1 of small salmon showed the proportion of 1 SW salmon were on average about $94 \%(\mathrm{n}=4757)$. In SFA 2 the 1SW component was on average about $97 \%(n=8872)$ of small salmon. There were no samples from commercial sampling in SFA 14B. In 1997, aged commercial samples indicated that the previous range was acceptable.

Total river returns of 2SW salmon (TRR) were calculated as follows:
TRR = LR2SW / (1-us)

The total river returns of small salmon are also calculated by equation 6 but from SR.
Spawning escapement (SE) or spawners was calculated according to the formula:
SE = TRR - AC,
where:
$\mathrm{AC}=$ angling catch which includes retained catch plus $10 \%$ of catch \& released mortality for released salmon.

A couple of modifications were made to the estimation procedure for Labrador in 1997. First, determination of exploitation rates was calculated separately for SFA 1, 2 and 14B using the active effort individually for each SFA. For SFA 1, the active number of licences declined from 141 in 1991 to 39 in 1997. For SFA 2, the active number of licences declined from 320 in 1991 to 99 in 1997. For SFA 14B, active licences declined from 52 in 1991 to 0 in 1997 when the fishery was closed. Exploitation rates determined as in equations 2, 3 and 4 were: SFA 1 - small was 0.0735 to 0.1399 and - large was 0.2221 to 0.3959 ; and SFA 2 - small was 0.0384 to 0.0728 and-large was 0.1589 to 0.2799 .

Numbers of small and large salmon for SFAs $1 \& 2$ were estimated from the exploitation model while for SFA 14B the results of assessments on Forteau Brook and Pinware River were expanded to inclade all the watersheds in SFA 14B. Returns to SFA 14B were 663 to 1545 small salmon and 146 to 327 large salmon.

Total mortalities of small and large salmon were accounted for by summing commercial catches of small salmon in Labrador and Newfoundland, large salmon in Labrador, Newfoundland, and Greenland, angling catches in Labrador of small and large salmon including $10 \%$ of the caught and released salmon, and small and large spawners. All of the above mortality estimates except catches of Labrador salmon in Newfoundland, 1969-1991 and Greenland could be obtained from equations 1 to 7. Catches in Newfoundland and Greenland were assessed as follows:

Greenland: for 1969-1992 and 1995-2004, removals of Labrador salmon by the Greenland fishery were assessed from data based on the sampling program in commercial fish plants at West Greenland (Anon, 1996). The Greenland fishery catches salmon that would have returned to homewaters as large salmon in the year following
e Greentand fishery. Numbers of Labrador salmon were determined by converting
atches in kg to numbers of salmon of 1SW North American origin that were of river age 4 and older. The number of Labrador salmon was estimated by assuming that $70 \%$ of the production of 4 -year and older river age salmon are from Labrador (Anon, 1993b).

Newfoundland: for 1969-1991, catches of Labrador small and large salmon in Newfoundland were included in total mortalities as the product of the ratio of tags caught in Newfoundland to Labrador and the catch in Labrador. For small salmon the ratio was $(24 /(100-24))=0.32$ and for large salmon it was $(41 /(137-41)=0.43$.

1998-2001
For the years, 1998-2001 when only one or two counting projects took place in Labrador, the raising factors of 1.04 to 1.49 for small salmon and 1.05 to 1.27 for large
salmon were used to estimate returns and spawners for Labrador from the overall PFA minus catches in Greenland, as was the case in previous years. However, in this case returns to rivers were derived for Labrador by subtracting landings in food fisheries. Also, catches in 1994-2006 were updated to reflect changes made to catch statistics in Labrador from the Licence Stub Return System. Procedures for the collection and compilation of commercial and angling fishery data are described in Ash and O'Connell (1987) for fishery years 1974-1996. For years 1969-1974, commercial catch data came from Anon. (1978). In 1997, the angling catch statistics were converted to a Licence Stub System (O'Connell et al., 1998) which continues to the present day.

## 2002-present

Counting projects occur on three to four Labrador rivers; out of about 100 extant salmon rivers. Because they occur on the same rivers each year, it is possible to extrapolate from abundance for small and large salmon per accessible drainage areas in these monitored rivers to unsurveyed ones in the remainder of Labrador. The accessible drainage areas were $9267 \mathrm{~km}^{2}$ for Lake Melville (SFA 1A), $25485 \mathrm{~km}^{2}$ for Northern Labrador (SFA 1B), $28160 \mathrm{~km}^{2}$ for Southern Labrador (SFA 2), and $2651 \mathrm{~km}^{2}$ for the Straits Area (SFA 14B). Accessible drainage area in the counting facility rivers was $1878 \mathrm{~km}^{2}$ resulting in an expansion factor of 35 to one. Not all rivers in Lake Melville were included due to a lack of information on presence of salmon populations in rivers in this region of Labrador. Lake Melville rivers whose drainage areas were included are Sebaskachu, Cape Caribou, Goose, MacKenzie, Kenamu, Caroline and Traverspine.

Abundances for SFAs 1A and 1B were derived from English River returns with maximum and minimum values developed using the observed variability of relative abundances in SFA 2. Total returns and spawners for Labrador are estimated by Monte Carlo simulation based on 10000 random draws from the range of values assuming abundances per $\mathrm{km}^{2}$ of accessible drainage were uniformly distributed. The relative abundances (per $\mathrm{km}^{2}$ ) for each SFA were then multiplied by the total accessible drainage area to derive total returns of small and large salmon. Ranges of values were developed to convert numbers of small and large salmon to numbers of 1SW and 2SW salmon from scale age information collected from counting fences and angling fisheries in Labrador. A bootstrap procedure was used to develop estimates of the proportions f sea age 1 salmon in estimates of small salmon returns and spawners, proportions of sea age 2 salmon in estimates of large salmon returns and spawners and proportions a age 1 salmon in the estimates of large salmon returns.

Sea age correction factors were:
Small to 1SW - 96 to $100 \%$
Large to 2SW - 60 to $71 \%$
Small overlap in large - 12 to $21 \%$

Spawners of 1SW and 2SW salmon were derived by subtraction of angling catches including an estimate of catch and release mortalities (10\%) from the returns.

### 2.3.1.2 Newfoundland

Inputs for the run-reconstruction model for Newfoundland include estimates of small, large and 2SW returns and spawners to rivers (minimum and maximum). The methods used to estimate returns and spawners to the rivers in Newfoundland are described by Reddin and Veinott (2010). In brief, returns and spawner estimates were derived from recreational fishery exploitation rates of retained small salmon for rivers with enumeration facilities; and ratios of large to small salmon were utilized to estimate large salmon. Exploitation rates were then applied to all rivers with reported angling catches. A non-parametric bootstrap technique was used, whereby exploitation rates and ratios of large to small salmon from rivers with enumeration facilities were chosen at random with replacement. The 95th confidence interval from 500 iterations of the weighted exploitation rate and ratio of large to small salmon was applied to angling catches on a Salmon Fishing Area (SFA) basis. The midpoint of the 95 th confidence interval was used as the minimum and maximum estimate returns of large and small salmon in each SFA. Estimates of 2SW returns are based on the expected proportion of 2 SW in the large salmon category ( $\geq 63 \mathrm{~cm}$ ). Commercial and recreational angling catches were derived as described for Labrador (2.3.1.1). Spawners in all years were determined as the returns to rivers minus angling catches including an adjustment for catch and release mortality.

### 2.3.1.3 Québec

In order to estimate abundance of stocks, rive C6) depending on the information available to estimate salmon returns (according to the method of Caron and Fontaine, 1999), with C1 being the most reliable evaluation and C6 the least. C1 corresponds to a river where the evaluation of the returns is based on a counting method, either from a fence or from a visual count through snorkelling or from a canoe. C 2 uses the same evaluation, but without knowing the number of small and large salmon, which is then estimated from proportions reported in the sport fishing landings and, if necessary, the catch and release. Salmon returns on C 3 rivers are determined based on multiple correlation factors, using catch number, fishing effort, season duration and river accessibility distance (Guillouët, 1993).

When estimation of the returns using a C1-C3 category is not possible, and when data of returns from previous years are available, the C4 category is used. C4 assumes that
terannual variations in salmon returns in the targeted river are approximately the same as variations observed in the other rivers of the corresponding region. Category C is for rivers where only landings data are available. In these rivers the salmon run is estimated from the average regional exploitation rate. Finally, a few small rivers have essentially no available data. C6 then assumes that the run is related to the available river salmon habitat and is estimated with respect to rivers of the same area for which run estimates and salmon habitat area are known. Estimated numbers of returns from C 4 to C 6 cannot be used to assess relative to attainment of conservation limits. However, they provide at least approximate numbers to estimate returns and spawners for salmon rivers in Québec.

The evaluation of the uncertainty associated with return estimates depends on the river category. For C 1 and C 2 rivers, the correction factor for the minimum and maximum number of returns is $+5 \%$ and $+10 \%$ for all rivers with a fish ladder and for all others in
zones Q1 to Q3 and Q10. The correction factor for rivers with darker water from zones Q5, Q6 and Q7 is $+10 \%$ and $+30 \%$. For the other categories, an uncertainty of $\pm 25 \%$ is associated with salmon return estimates, except for category C3 where calculation depends on the method of Guillouët (1993).

The number of spawners is obtained using the return estimate minus all river catches, which include landings and other types of removal. In most cases, river catches include landings from sport fishing only, which may be conducted by indigenous people such as that on the Betsiamites River. The other types of removal are of limited number and include mainly natural mortality, salmon captured for hatchery use and subsistence fishing when practised in river.

Overall return estimates for all Québec rivers are obtained by adding in-river salmon returns, commercial fishing (when operated), indigenous people subsistence fishing when practised in estuaries and an estimate of non-reported landings. Howeyer, little scientific data are available on non-reported landings and thus, estimates are based on good judgment, following consultations with regional biole

### 2.3.1. 4 Gulf

Estimation of returns and spawners are developed for the four salmon fishing areas of Gulf Region (SFAs 15 to 18).

## SFA 15

The major river in this area is the Restigouche River. The returns and spawners are estimated for the Restigouche River exclusive of returns to the Matapedia River, which are included in Québec zone Q1. The Restigouche River stock assessment is based on angling catch with assumed exploitation rates between $30 \%$ and $50 \%$ with estuary catches added back after the estimates of returns. Return and spawner estimates for SFA 15 are based on Restigouche River data, scaled up for SFA 15 using angling data. The return and spawner estimates for SFA 15 are derived from the return and spawner estimates for Restigouche(New Brunswick). The minimum and maximum return and spawner estimates are derived from the minimum and maximum ratios of angling catch in all of SFA15 relative to angling catch in Restigouche (New Brunswick) (min = $1.117 ; \max =1.465$ ). Harvests represent retained angling catch plus $6 \%$ catch and release mortality for released fish. The proportion of 2SW in large salmon numbers is based on aged scale samples from angling, trapnets, and broodstock. In the years when no cale samples analysis is available, a mean value of 0.65 is used.

SFA 16
The most important Atlantic salmon river in SFA 16 is the Miramichi River. The Miramichi makes up $91 \%$ of total rearing area of SFA 16 and returns to the river are assessed annually. For 1971 to 1991, minimum and maximum values are based on capture efficiencies of the Millbank estuary trapnet representing a lower CI of $-20 \%$ of the estimate and upper CI of $33 \%$ of the estimate. For 1992 to 1997 , minimum and maximum are lower and upper CI and based on estimate bounds of $-18.5 \%$ to $+18.5 \%$. Since 1998 to the present, minimum and maximum are 5th and 95th percentile range from a Bayesian hierarchical model used in the assessment. Returns to SFA 16 are Miramichi returns (Minimum, Maximum) / 0.91. Proportion 1SW in small salmon is
from scale ageing; proportions have varied from 0.97 to 1.0 . Proportion 2 SW in the large salmon category is obtained from scale ageing. Spawners are returns minus harvests. For 1998 to 2018, the harvest of large salmon is estimated as the sum of the indigenous fisheries harvests for large salmon and $1 \%$ of the large salmon catch ( $30 \%$ exploitation rate, $3 \%$ catch and release mortality). Prior to 1995 , the harvest of small salmon is estimated as $30 \%$ of the small salmon return plus the harvest from the indigenous fisheries. During 2015 to 2018, when mandatory catch and release management measures were in effect, the fisheries related losses (harvests) of small salmon were estimated as the sum of the indigenous fisheries harvests for small salmon and $1 \%$ of the small salmon catch ( $30 \%$ exploitation rate, $3 \%$ catch and release mortality).

## SFA 17

For 1970-1994, small returns are estimated from retained small salmon catch in the Morell River divided by the river-specific exploitation rate. Salmon catch in the Morell River was estimated in 1970-1990 by DFO Fisheries Officers, and in 1991, 1992, and 1994 by angler mail-out surveys. The number of small retained salmon in 1993 was not recorded, so the number used is the mean for 1986-1992. For 1970-1993, exploitation rate was taken as the mean of exploitation rates estimated for 1994, 1995, and 1996 (0.317). For 1994, exploitation rate was 0.34 . The min and max of small returns are calculated using exploitation $+/-0.1$; e.g. $0.34+/-0.1$ gives 0.24 and 0.44 . Large returns $=$ (number of small returns/proportion small) - number of small returns. For 1970-1980, proportion small is calculated from numbers of small and large salmon in the angling catch of each year. For 1981-1994, proportion small is taken from counts at the Leards Pond trap on the Morell River. Small spanpers = number of small recruits - number of small retained. Large spawners = number of large recruits - number of large retained. In 2012, the Province of Prince Edward Island discontinued the sale of recreational fishing licences for Atlantic salmon. Instead, anglers who purchased a trout licence are authorized to also fish for Atlantic salmon. Since it was no longer possible to assemble a list of salmon anglers, the salmon angler survey was discontinued from 2012. In the absence of salmon angling data for 2012 and subsequently, catch statistics estimated for 2011 are used for 2012 and subsequent years.

Spawner estimates for 1995 to the present are derived from redd counts in 23 rivers. Foryears and rivers in which redd counts are unavailable, redd numbers are estimated by linear interpolation from the preceding and succeeding count year. Redd numbers ears prior to the first count are taken as the first count. Redd numbers in years after the last count are taken as the last count. Female spawners are estimated from the ratio of 3.357 redds/female spawner, measured in the West River in 1990. Total spawners are estimated from size-specific sex ratios derived from counts at Leards and Mooneys Ponds, Morell River, in 1986-2001. The proportion of large salmon is assumed to be 0.5 in the Cains, Carruthers, Trout (Coleman), Morell, Cardigan, West, and Dunk Rivers, and 0.9 in all other rivers. Spawners are presented as Min (estimated spawners -20\%) and Max (estimated spawners $+20 \%$ ). Returns are spawners + total estimated fishing mortality, including angler catches, catch and release mortality, and indigenous harvests. Angler catches and catch and release mortality are estimated from angler card surveys. Returns are presented as Min (estimated returns - 20\%) and Max (estimated returns $+20 \%$ ). It is assumed that large salmon and 2 SW salmon are equivalent.

SFA 18
Returns and spawners to SFA 18 are derived from estimates of returns and spawners to the Margaree River, adjusted for the ratio of the SFA 18 angling catch to the Margaree River catch. For small salmon, the ration of SFA 18 catch to Margaree catch varies between 1.15 and 2.71 for years 1984 to 2004. For large salmon, the ratio of SFA 18 catch to Margaree catch varies between 1.08 and 2.32 for years 1984 to 2004. Returns to Margaree River are estimated using various techniques.

- 1970 to 1983 angling catch divided by range of exploitation rates with maximum exploitation rate of 0.37 and minimum exploitation rate of 0.215 ;
- 1984 to 1986 based on annual assessments;
- 1987 to present angling catch and effort data from logbooks and provincial licence stubs are used to derive the returns. The catchability coefficient per rod day is estimated from angling catch and effort data for the years 1988 to 1996 when mark and recapture programmes were used to estimate returns, independently from angling data.

Spawners for 1970-1983 equal returns minus removals. Spawners for 1984 to the present equal returns minus catch for small salmon and returns minus catch, corrected for $5 \%$ mortality, for large salmon. 2SW salmon represent between 0.77 and 0.87 of large salmon returns and spawners.

### 2.3.1.5 Scotia-Fundy

Salmon originating in rivers of the Atlantic coast of Nova Scotia and southwest New Brunswick in Salmon Fishing Areas (SFAs) 19-21 and the portion of SFA 23 outside the inner Bay of Fundy comprise the Scotia-Fundy stocks. With the exception of at least one stock in SFA 19, they have a large salmon component that migrates to the North Atlantic/Labrador Sea (Amiro et al., 2008). Estimates of returns and spawning escapement for the Scotia-Fundy stocks are provided as inputs to the runreconstruction model. Inner Bay of Fundy Atlantic Salmon (SFA 22 and part of SFA 23) have been federally listed as endangered under the Canadian Species at Risk Act and uded as inputs into the run-reconstruction model. With the exception of one population, inner Bay of Fundy stocks have a localized migration strategy while at sea and an incidence of maturity after one winter at sea.

Consistent with the requirements of the model, a range (minimum to maximum) of returns and spawning escapement for the Scotia-Fundy stocks is provided for the runreconstruction model. The methods used to estimate total returns and spawners are described by Amiro et al. (2008). In brief, for SFAs 19-21, the escapement is based on the count of small and large salmon at the Morgan Falls fish-way on the LaHave River from 1970 to the present year, scaled up to the region using the relationship between this count and the recreational catch data for rivers in SFA 19 to 21 from 1970 to 1997 and a catch rate for the LaHave River from 1970 to 1997. Estimates of the returns also include estimates of landings in the commercial salmon fisheries in SFA 19-21 from 1970 to 1983. The model is fitted using maximum likelihood, and the $90 \%$ confidence limits are carried forward as the minimum and maximum values. In SFA 23 from 1970 until 1992, estimates of total 1SW and large wild-origin salmon returns are based on the estimated number of returns destined for tributaries above Mactaquac Dam on the

Saint John River; this includes in-river and outer-Fundy commercial landings (19701971 and 1981-1983), in-river indigenous harvests (since 1974), and counts at Mactaquac Dam. These estimates are raised by the proportion of the total accessible productive habitat in SFA 23 that is upstream of Mactaquac Dam (0.4-0.6). Hatcheryorigin returns were attributed to above Mactaquac Dam only and no hatchery 1SW and MSW returns were estimated for other rivers within SFA 23 (outer Fundy). Since 1993 the estimates of 1SW and MSW returns to the Nashwaak River have been used to estimate the wild production from tributaries of the Saint John River below Mactaquac Dam. The estimated 1SW and MSW returns to the Nashwaak River (above Counting Fence), is raised by the proportion of the total production area accounted for below Mactaquac ( $0.21-0.3$ ) and then added to the above Mactaquac totals.

### 2.3.1.6 USA

Total salmon returns and spawners for USA rivers are based on trap and weir catches and for the small rivers in Maine that do not have fish counting facilities, estimates of spawners were based on redd counts.

### 2.3.2 Improvements to NAC input data

Modifications to input variables used in assessments for the NAC area are reported by WGNAS in the year in which they are first implemented.

Over recent years, efforts have been made toreduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes). However, the methods used to derive estimates of unreported catch vary markedly between countries. For example, some countries include only illegally caught fish in the unreported catch, while other countries include estimates of unreported catch by legal gear as well as illegal catches in their estimate

Descriptions of the national approaches used for evaluating unreported catches have been reported at various WGNAS meetings (e.g. ICES, 1996; 2000; 2002; 2010a). In addition, detailed reports describing national procedures for evaluating illegal and unreported catch, and efforts to minimise this, were submitted by parties to NASCO in 2007 in support of a special theme session on this issue. Full details are available at: http:/fww

Input data commonly rely on rod catches and the practice of catch and release has become increasingly important in recent years to reduce levels of exploitation on stocks. As the practice is increasing, WGNAS have previously recommended (ICES, 2010a) that consideration should be given to incorporating mortality associated with this practice in river-specific, regional and national assessments.

The procedures currently used to incorporate catch and release and unreported catches into regional, national and international assessments are summarised at Appendix 1 (from ICES, 2010a).

### 2.4 Biological and other data requirements

As noted previously, many of the 'conventional' data requirements (e.g. marine survey data and commercial CPUE) used in the assessment of other commercially important
fish species are inappropriate to salmon. A range of biological, catch and exploitation rates and other data pertinent to appropriate stock assessments are however, collected and made available to WGNAS to help inform assessments and to aid in responding to the various questions posed by NASCO.

Appendix 2 of this Stock Annex provides an overview of current and possible future data requirements for Atlantic salmon assessment/ scientific advice. This was compiled at a recent meeting of WGNAS (ICES, 2013) in relation to monitoring requirements under the European Data Collection Framework (DCF) and following a more detailed review of the data requirements under DCF (ICES, 2012c). This table illustrates the type of information collected/available, but is provided for illustrative purposes only. It should be noted that many Atlantic salmon producing countries fall outside the DCF provisions, which only relate to countries within the European Union. Further, Sovereign states are responsible for the regulation of salmon fisheries within their areas of jurisdiction. Formal ICES catch advice is only required for the distant water salmon fisheries, which take salmon originating in rivers of another party.

3 Assessment methods

In managing Atlantic salmon fisheries, NASCO has adopted a fixed escapement strategy (Potter, 2001), in recognition of the importance of the spawning stock to subsequent recruitment. Therefore, in managing the distant water fisheries at Faroes and West Greenland, the spawning requirements, of the rivers contributing to these fisheries must be defined. Management advice, expressed as allowable harvest (tonnes), is then predicated on a forecast of salmon abundance prior to the fishery such that the spawning requirements of the contributing stocks can be achieved. The provision of catch advice thus proceeds through a number of steps:

The definition of spawning objectives;
he development of a measure of abundance prior to the fishery; i.e. the prefishery abundance or PFA;
measure of the spawning stock contributing to the PFA;
A model to forecast the PFA;
The development of a risk analysis framework for the catch advice.

These steps are described in detail in the following sections, subdivided as necessary for the different distant water fisheries and the various stock complexes which contribute to the two fisheries (Greenland and Faroes).

### 3.1 Definition of spawning objectives

### 3.1.1 Management objectives and reference points

Conservation limits (CLs) for North Atlantic salmon have been defined by ICES as the stock level that will achieve long-term average maximum sustainable yield (MSY). NASCO has adopted the following definition of CLs (NASCO, 1998): ‘The CL is a limit
reference point; having populations fall below these limits should be avoided with high probability.'

Atlantic salmon has characteristics of short-lived fish stocks; mature abundance is sensitive to annual recruitment because there are only a few age groups in the adult spawning stock. Incoming recruitment is often the main component of the fishable stock. For such fish stocks, the ICES MSY approach is aimed at achieving a target escapement (MSY Bescapement, the amount of biomass left to spawn). No catch should be allowed unless there is a high probabaility that this escapement can be achieved. The escapement level should be set so there is a low risk of future recruitment being impaired, similar to the basis for estimating $B_{p a}$ in the precautionary approach. In shortlived stocks, where most of the annual surplus production is from recruitment (not growth), MSY Bescapement and $B_{p a}$ might be expected to be similar.

It should be noted that this is equivalent to the ICES precautionary target reference points $\left(S_{p a}\right)$. Therefore, stocks are regarded by ICES as being at full reproductive capacity only if they are above the precautionary target reference point This approach parallels the use of precautionary reference points used for the provision of catch advice for other fish stocks in the ICES area.

For the assessment of the status of stocks and advice on management of national components and geographical groupings of the stock complexes, where there are no specific management objectives:

- ICES requires that the lower bound of the $90 \%$ confidence interval of the current estimate of spawners is above the CL for the stock to be considered at full reproductive capacity (equivalent to a probability of at least $95 \%$ of meeting the CL).
- When the lower bound of the confidence limit is below the CL, but the midpoint is above, then ICES considers the stock to be at risk of suffering reduced reproductive capacity.

Finally, when the midpoint is below the CL, ICES considers the stock to be suffering reduced reproductive capacity.

Ideally, Atlantic salmon should be assessed and managed on the basis of river-specific stock units, the scale corresponding best to the spawner to recruitment dynamic Chaput, 2012). In reality, this is not the case for the majority of rivers, although efforts continuing to develop river-specific CLs and assessment protocols and developments are reported annually to WGNAS (e.g. ICES, 2013).

The risk assessment frameworks applied by WGNAS directly evaluate the risk of meeting or exceeding the stock complex objectives. Managers can choose the risk level which they consider appropriate. ICES considers however that to be consistent with the MSY and the precautionary approach, and given that the CLs are considered to be limit reference points and to be avoided with a high probability, then managers should choose a risk level that results in a low chance of failing to meet the CLs. ICES recommends that the probability of meeting or exceeding CLs for individual stocks should be greater than $95 \%$ (ICES, 2012b).

### 3.1.2 Reference points in the NEAC area

River-specific CLs have been derived for salmon stocks in some countries in the NEAC area (France, Ireland, UK (England \& Wales), Norway and Sweden). An interim approach has been developed for estimating national CLs for countries that cannot provide one based upon river-specific estimates. This approach is based on the establishment of pseudo stock-recruitment relationships for national salmon stocks (Potter et al., 2004).

The NEAC-PFA run reconstruction model (see below) provides a means of relating estimates of the numbers of recruits to the numbers of spawners. The numbers of 1SW and MSW spawners are converted into numbers of eggs deposited using the proportion of female fish in each age class and the average number of eggs produced per female. The egg deposition in year ' $n$ ' is assumed to contribute to the recruitment in years " $n+3$ " to " $n+8$ " in proportion to the numbers of smolts produced of ages 1 to 6 years respectively. These proportions are then used to estimate the lagged egg deposition' contributing to the recruitment of maturing and non-maturing 1SW fish in the appropriate years. The plots of lagged eggs (stock) against the 15W adults in the sea (recruits) are presented as 'pseudo stock-recruitment' relationships for each homewater country or region that is unable to provide river-specific CLs. In countries where with more than one region, the analysis is carried out for each region separately and the resulting estimates are summed to provide a national figure.

As noted previously, ICES currently define the CL for salmon as the stock size that will result in the maximum sustainable yield (MSY) in the long term. However, it is not straightforward to estimate this point on the stock-recruitment relationships established by the national PFA run-reconstruction models, as the replacement line (i.e. the line on which'stock' equals'recruits') is not known for these relationships. This is because the stock is expressed as eggs, while the recruits are expressed as adult salmon. To address this, WGNAS has developed a method for setting biological reference points from the national/ regional pseudo stock-recruitment datasets (ICES, 2001). This model assumes that there is a critical spawning stock level below which recruitment decreases linearly towards zero and above which recruitment remains constant. The position of this critical stock level is determined by searching for the stock value that proyides the line of best fit for the stock and recruitment data provided by the PFA run-reconstruction model as determined by the residual sum of squares. This point is a proxy for $S_{\text {lim }}$ and is therefore defined as the CL for the stock, and is indicated by the flection point in the hockey-stick relationship (e.g. see example at Figure 3.1.2.1).


Figure 3.1.2.1. Pseudo stock-recruitment relationship for UK (Scotland) eastern region (from ICES, 2013).

Where river-specific estimates of CLs have been derived for all the rivers in country or region, these are aggregated to provide national estimates. For countries where the development of river-specific CLs has not been completed, the method described above has been used (see example in Table 3.1.2.1, from ICES, 2013). The estimated national CLs are then summed to provide aggregate CLs for the northern and southern NEAC stock complexes (Table 3.1.2.1).

The CLs have also been used to estimate the spawning escapement reserves (SERs). These represent the CLs increased to take account of natural mortality between the recruitment date, 1st January, and the return to homewaters for maturing and nonmaturing 1SW salmon from the northern NEAC and southern NEAC stock complexes (Table 3.1.2.1).

Table 3.1.2.1. Conservation limits (CLs) for NEAC countries and stock complexes estimated from river-specific yalues, where available, or the national PFA run-reconstruction model. Spawner escapement reserves (SERs) are also included for each stock complex (ICES 2016).


WGNAS considers the current CL and SER levels may be less appropriate to evaluating the historical status of stocks (e.g. pre-1985), that in many cases have been estimated with less precision.

### 3.1.3 Reference points in the NAC area

In many regions of North America, the CLs are calculated as the number of spawners required to fully seed the wetted area of the river. The methods and values used to derive the egg and spawner conservation requirements for Atlantic Canada are documented in O'Connell et al. (1997). CLs have generally been derived using freshwater production dynamics translated to adult returns to estimate the spawning stock for maximum sustainable yield (MSY). Data were available for a limited number of stocks and these values were transported to the remaining rivers using information on habitat area and the age composition of the spawners. A similar procedure was used to determine the CLs for rivers in the USA (ICES, 1995). In Québec, adult-to-adult stock-recruitment relationships for six rivers were used to define the CLs for the other rivers (Caron et al., 1999). The definition of conservation in Canada varies by region and in some areas, historically the values used were equivalent to maxinizing /optimizing freshwater production. These are used in Canada as limit reference points and they do not correspond to MSY values. Reference points for Atlantic salmon that conform to the Precautionary Approach have been recently reviewed in eastern Canada (DFO 2015).

The NAC conservation requirements for $2 S W$ salmon (only these are required in developing catch options for the West Greenland fishery) are summarised in Table 3.1.3.1 (from ICES, 2013). These are calculated from the adult age structure within the different regions and total 123349 2SW salmon for Canada and 29199 2SW salmon for the USA, for a combined total of 152548 .

Table 3.1.3.1. 2SW Conservation limits (CLs) for the six regions in the NAC area estimated from river-specific values.

In 2016, WGNAS provided an overview of ongoing work that Fisheries and Oceans Canada (DFO) was undertaking to refine reference points for Atlantic salmon in Canada that conformed to the Precautionary Approach (ICES 2016a). DFO Maritimes Region (Scotia-Fundy) retained the current conservation requirement based on 240 eggs per $100 \mathrm{~m}^{2}$ as the Limit Reference Point (DFO 2012; Gibson and Claytor 2013). DFO Newfoundland Region retained the current conservation requirement based on 240 eggs per $100 \mathrm{~m}^{2}$ of fluvial rearing habitat, and in addition for insular Newfoundland 368 eggs per ha of lacustrine habitat (or 150 eggs per ha for stocks on the northern peninsula of Newfoundland), as equivalent to their Limit Reference Point
and have defined the Upper Stock Reference as $150 \%$ of the Limit Reference Point (DFO 2017). The Province of Quebec revised the Limit Reference point and Upper Stock Reference point using a Bayesian hierarchical analysis of stock-recruitment data (Dionne et al., 2015; ICES 2017), and DFO Gulf Region undertook an exercise in 2017 to revise and define the Limit Reference Point in that region of Canada using the proportion of eggs from MSW salmon as a covariate in the Bayesian Hierarchical Model (DFO 2018). The Limit Reference Points in all cases are defined in terms of total eggs from all sizes and sea ages of salmon. Revised reference points specific to 2SW salmon for DFO Gulf Region and Quebec were presented at the working group meeting in 2019. The revised 2 SW conservation limit ( 18737 fish) for Gulf is a $38 \%$ decrease from the previous value whereas the revised 2SW CL for Quebec is a slight increase (9\%) from the previous value. Revised CLs for 2SW salmon for Canada total 114295 (previous value 123 349) and 29199 for the USA, for acombined revised total of 143494 2SW salmon (previous value 152548 ). The revised 2 SW CL values will be used in the next full assessment and catch advice period, in 2021, or earlier if the framework of indicators for 2020 indicates that an earlier assessment and revised catch advice would be warranted.

No other changes to the 2SW CLs or the Management Objectives were made from those identified previously (ICES, 2015).

### 3.2 Estimating PFA

Estimates of PFA are derived by run-reconstruction methods. These work back in time from estimates of abundance in homewaters to earlier periods of the salmon's life cycle by adding in catches at appropriate times and adjusting for survival. The runreconstruction approach was first presented at ICES in 1992 and was subsequently adopted for stocks on both sides of the Atlantic (Rago et al., 1993a; Potter and Dunkley, 1993; Potter et al, 1998; 2004). The main advantage of backwards-running, runreconstruction models over alternative forward-running approaches is that more extensive data are available on adult returns (e.g. traps, counters and catch data) than on freshwater production of juveniles. In addition, rates of natural mortality (M) were thought to be lower and more stable for large salmon after their first winter in the sea than during the post-smolt phase (Potter et al., 2003).
The models used to estimate PFA take the generalised form:

$$
P F A=N h * \exp \left(M t_{h}\right)+\sum_{i} C_{i} * \exp \left(M t_{i}\right)
$$

Where: Nh is the number of adult fish returning to homewaters, $C_{i}$ the catch of fish from the stock in each interception fishery i (operating before the fish return to homewaters), M the monthly instantaneous rate of natural mortality of salmon in the sea after the first sea-winter, $t_{i}$ the time in months between the PFA date and the midpoint of fishery $i$, and $t_{h}$ is the time in months between the PFA date and the midpoint of the return of fish to homewaters. Coastal catches are also added to the estimate where appropriate.

### 3.2.1 NEAC area run reconstruction model

The original model used to estimate the PFA of salmon from countries in the NEAC area was described by Potter et al. (2004); modifications have been described in subsequent WGNAS reports. PFA in the NEAC area is defined as the number of 1SW recruits on January 1st in the first sea winter. As there are relatively few fish of sea age three or more in most stocks, the model caters for two age groups, 1SW and MSW, the latter including all fish of sea age two or more that are treated as a single cohort. The model is therefore based on the annual catch in numbers of 1SW and MSW salmon in each country. These are raised to take account of minimum and maximum estimates of non-reported catches and exploitation rates of these two sea-age groups.

Thus, for each country (or region) c in year $y$, the total number of fish of sea age a caught in homewater fisheries $\left(\mathrm{Ch}_{a, y, c}\right)$ is calculated by dividing the declared catch ( $\mathrm{Cd}_{\mathrm{a}, \mathrm{y}, \mathrm{c}}$ ) by the non-reporting rate ( $1-\mathrm{U}_{\mathrm{a}, \mathrm{y}, \mathrm{c}}$ ):

$$
\mathrm{Ch}_{\mathrm{a}, \mathrm{y}, \mathrm{c}}=\mathrm{Cd}_{\mathrm{a}, \mathrm{y}, \mathrm{c}} /\left(1-\mathrm{U}_{\mathrm{a}, \mathrm{y}, \mathrm{c}}\right)
$$

where: $\mathrm{U}_{\mathrm{a}, \mathrm{y}, \mathrm{c}}$ is the estimated proportion of the total catch that is unreported or discarded. The number of fish returning to homewaters ( $\mathrm{Nha}, \mathrm{y}, \mathrm{c}$ ) is estimated by dividing the total homewater catch by the exploitation rate $\left(\mathrm{H}_{\mathrm{a}, \mathrm{y}, \mathrm{c}}\right)$ :

$$
N h_{a, y, c}=C h_{a, y, c} / H_{a, y, c}
$$



As the model provides estimates of total returns and total catch (including non-catch fishing mortality), it is then also possible to estimate the spawner escapement ( $\mathrm{Ns}_{\mathrm{a}, \mathrm{y}, \mathrm{c})}$ :


Total catches in the Faroese $\left(\mathrm{Cf}_{\mathrm{a}, \mathrm{y}}\right)$ and West Greenland $\left(\mathrm{Cg}_{\mathrm{a}, \mathrm{y}}\right)$ fisheries are similarly calculated by correcting the declared catches for non-reporting, but they are not raised for the exploitation rate, because the uncaught fish are accounted for from the returns to homewaters. The West Greenland fishery only exploits salmon that would otherwise MSW fish, although the majority are 1SW fish in the summer that they are caught; for the purpose of the model, all are classed as 1SW. The Faroese fishery exploits predominantly MSW salmon, but also a small number of 1SW fish, $78 \%$ of hich have been estimated to be maturing (ICES, 1994). The Faroese fishery has not taken salmon since 2000, but over the two decades previous to that, a substantial proportion of the fish caught in the Faroese fishery were escapees from salmon farms, and these are discounted from the assessment of wild stocks on the basis of data from Hansen et al. (1999). The incidence of farm escapees in the West Greenland catch is thought to be $<1.5 \%$ (Hansen et al., 1997), so this portion is ignored in the model. The total estimated catches of wild fish in both distant-water fisheries are assigned to the PFA for different countries on the basis of historic tagging studies (Potter, 1996).

The returns to homewaters and catches in the distant-water fisheries of 1SW and MSW salmon are then raised to take account of the marine mortality between January 1st in the first sea winter (the PFA date) and the mid-point of the period over which the respective national fisheries operate. WGNAS determined a natural mortality value of 0.03 per month to be the most appropriate (ICES, 2002) and a range 0.02 to 0.04 is
applied within the model in a Monte Carlo simulation. Thus, the PFA of maturing 1SW fish (PFAm), survivors of which will return to homewaters as 1SW adults, is:

$$
\operatorname{PFAm}_{y, c}=\mathrm{Nh}_{1, y, c}{ }^{*} \exp \left(\mathrm{Mt}_{\mathrm{t}, 1, \mathrm{c}}\right)+0.78{ }^{*} \mathrm{Cf}_{1, \mathrm{y}}{ }^{*} \mathrm{w}_{y}{ }^{*} \mathrm{pf}_{1, \mathrm{c}}{ }^{*} \exp \left(\mathrm{Mtf}_{\mathrm{f}, 1, \mathrm{c}}\right)
$$

and the PFA of non-maturing 1SW fish (PFAn), survivors of which will return to homewaters as MSW adults, is:

$$
\begin{aligned}
& \operatorname{PFAn} n_{y, c}=N h_{2, y+1, c} * \exp \left(\mathrm{Mt}_{\mathrm{t}, 2, \mathrm{c}}\right)+\mathrm{Cg}_{1, \mathrm{y}}{ }^{*} \operatorname{pg}_{1, \mathrm{c}}{ }^{*} \exp \left(\mathrm{Mt}_{\mathrm{g}, 1, \mathrm{c}}\right) \\
& +0.22{ }^{*} \mathrm{Cf}_{1, y}{ }^{*} \mathrm{w}_{y}{ }^{*} \mathrm{pf}_{1, c}{ }^{*} \exp \left(\mathrm{Mtf}_{\mathrm{f}, 1, \mathrm{c}}\right)+\mathrm{Cf}_{2, \mathrm{y}+1}{ }^{*} \mathrm{w}_{\mathrm{y}+1}{ }^{*} \mathrm{pf}_{2, c}{ }^{*} \exp \left(\mathrm{Mt}_{\mathrm{t}, 2, \mathrm{c}}\right)
\end{aligned}
$$

where indices y and c represent year and country/region, indices 1 and 2 the 1SW and MSW sea age groups, w is the proportion of the Faroese catch that is of wild origin, pf and pg are the proportion of the catches in the Faroese and West Greenland fisheries originating in each country (as indexed), and $t_{h}, t_{f}$ and $t_{g}$ are the times in months between the PFA date and the midpoints of the homewater fisheries, the Faroese fishery, and the West Greenland fishery, respectively, for the year classes and country/region as indexed.

Total 1SW recruitment for the NEAC area in yeary is therefore the sum of the maturing 1SW and non-maturing 1SW recruitments for that year for all countries:


The non-reporting rates, exploitation rates, natural mortality, and migration times in the above equations cannot be estimated precisely, so national experts provide minimum and maximum values based upon best available knowledge that are considered likely to be centred on the true values (ICES, 2003). A Monte Carlo simulation (MSC: 12000 trials) is used to estimate confidence intervals on the stock estimates.

Where appropriate to the provision of management advice, the national outputs from the model are combined into stock complexes, such as those for southern and northern NEAC (ICES, 2002). The confidence limits for these combined estimates are derived from the sum of the national variances obtained from the MCS (the covariances are sumed to be small). This model has provided time-series of PFA estimates for NEAC almon stocks from 1971 to the present.

The model was initially run using 'Crystal Ball' (CB) in Excel (Decisioneering, 1996). However, an updated version of the model which runs in the ' $R$ ' programming language (R Development Core Team, 2007) was developed in 2011 (ICES, 2011). This provided a more flexible platform for the further development of the model and to allow its integration with the Bayesian forecast model for the development of catch options (see below). In 2012, the outputs of the CB and ' $R$ ' models were compared to examine the approaches taken and to validate the outputs (ICES, 2012a). Since 2013, the run-reconstruction analysis has been completed by WGNAS using the ' R ' programme (ICES, 2013). This has also enabled additional sources of uncertainty to be incorporated into the modelling approach (ICES, 2013).

The full set of country-specific data inputs, as used in the most recent assessment (ICES, 2016) is provided at Appendix 3. The ' $R$ ' code used for running the model and the additional data input file required to run the model, are available on the ICES WGNAS SharePoint site.

### 3.2.2 NAC area run reconstruction model

The run-reconstruction model developed by Rago et al. (1993a) and described in previous WGNAS reports (ICES, 2008; 2009a) and in the primary literature (Chaput et al., 2005) is used to estimate returns and spawners by size (small salmon, large salmon) and sea age group (2SW salmon) to the six geographic regions of NAC from 1971 to the present. The model takes the form:

$$
P F A_{y \operatorname{ear}(\mathrm{i})}=\left[\mathrm{NR} 2_{y \operatorname{year}(\mathrm{i}+1)} * \mathrm{e}^{\mathrm{MX} 1}+\mathrm{NC} 2_{y \operatorname{year}(\mathrm{i}+1)}\right]^{*} \mathrm{e}^{\mathrm{MX} 10}+\mathrm{NC} 1_{\text {year }(\mathrm{i})}+\mathrm{NG} 1_{y} \text { yea }
$$

where: NR2 year( $(i+1)$ is the sum of 2 SW returns to six regions of North America in year $i+$ $1, \mathrm{NC} 2$ year $(i+1)$ is the catch of 2 SW salmon in Newfoundland and Labrador commercial fisheries in year $\mathrm{i}+1, \mathrm{NC1}_{\text {year( }(\mathrm{i})}$ is the catch of 1 SW non-mataring salmon in Newfoundland and Labrador commercial fisheries in year $\mathrm{i}, \mathrm{NG} \mathrm{l}_{\text {year }(\mathrm{i})}$ is the catch of 1SW non-maturing salmon of North American origin in the Greenland fishery in year $i$, and M is the monthly instantaneous natural mortality of 0.03 .

The reconstruction begins with the estimation of returns of 2 SW salmon in year $i+1$ to six regions in eastern North America: Labrador, Newfoundland, Québec, Gulf, ScotiaFundy, and USA. For the four southern regions, the regional returns include the harvest in the coastal commercial fisheries but this is not the case for Newfoundland and Labrador. For Labrador, the returnsto rivers are estimated from the commercial harvest factored by an exploitation rate. The harvest of 2 SW salmon in the Newfoundland and Labrador mixed-stock fisheries in year $i+1$ is added to the sum of the returns to the six regions (prorated backward for one month of natural mortality equates to 1st June of year $i+1$ ) to produce the returns to North America. Finally, the harvests of North American origin salmon in the Greenland fisheries in year i and the harvest of non-maturing 1SW salmon in the Newfoundland and Labrador commercial ries in year i are added to the prorated returns to North America (ten months between abundance at Greenland on 1st August year i and North America on 1st June year $i+1$ ) to produce the pre-fishery abundance of non-maturing 1SW salmon of North American origin. An instantaneous natural mortality rate of 0.03 per month is assumed salmon in the second year at sea for all years (ICES, 2002). Adjustments to the input data resulting from reductions and subsequent closures of commercial fisheries in North America are summarized by Friedland et al. (2003).

Following earlier WGNAS recommendations (ICES, 2008), the run-reconstruction model since 2009 has been developed using Monte Carlo simulation (OpenBUGS; http://mathstat.helsinki.fi/openbugs/; Lunn et al., 2000). This is similar to the approach applied for the NEAC area.

The PFA of the non-maturing component of 1SW fish, destined to be 2SW returns (excluding 3SW and previous spawners) is the estimated number of salmon in the North Atlantic on August 1st of the second summer at sea. As this requires estimates
of 2 SW returns to rivers, there is always a lag in providing this figure (PFA estimates for year $n$ require 2 SW returns to rivers in North America in year $n+1$ ).

The full set of data inputs, as used in the most recent assessment (ICES, 2015) is provided at Appendix 4. The ' $R$ ' code used for running the model is available on the WGNAS SharePoint site.

### 3.2.3 Instantaneous natural mortality rate (M)

The natural mortality rate for salmon after they recruit to the distant water fisheries has been the subject of much discussion. WGNAS originally used a value of 0.01 per month, based upon Doubleday et al. (1979), but this was modified to 0.03 per month following a detailed review as part of the EU SALMODEL project (Crozier et al., 2003; ICES, 2002) on the basis of inverse-weight and maturity-schedule models. The rate is assumed to have been constant over the time-series. While mortality may be expected to vary among years and may also be different for maturing and non-maturing 1SW recruits, WGNAS has not had data on which to base the use of different values, or values that change over time. The assumption is, therefore, that the mortality of adult fish after the first winter at sea has not changed and that all the variability of marine mortality has occurred at the post-smolt stage. Efforts are continuing to explore levels of mortality and to better partition this between different stages of the life cycle. The issue was also subject to further investigation within the EU ECOKNOWS project and Bayesian modelling may provide alternative approaches in future.

### 3.3 PFA forecast models

### 3.3.1 Introduction

The provision of quantitative catch advice for the distant water fisheries requires estimates of abund ance before the fisheries take place. While there has been some use of in-season surveys in the management of these fisheries (NASCO, 2001), such methods are considered too impractical and costly to implement on a widespread scale (Potter et al. 2004). Models have therefore been developed by WGNAS which relate abundance estimates obtained at other life stages to the PFA. The objective has been to account for this relationship in terms of biological or environmental factors that affect natural mortality, and to use this to forecast future stock levels.
n initial PFA forecast model for North American stocks (Rago et al., 1993b) utilised dices of thermal habitat in relation to historically observed PFA (from the runreconstruction model) to predict future PFA. Similar approaches were explored by Crozier et al. (2003) for NEAC stocks. However, while statistically significant temperature indices could be constructed, the relationships were not always consistent or intuitively correct. Alternative approaches were therefore explored for NEAC; these are described by Potter et al. (2004). More recently work by the ICES Study Group on Salmon Stock Assessment and Forecasting (SGSSAFE) has, however, resulted in the development of Bayesian forecast models for both NAC and NEAC (ICES, 2009a; 2011; Chaput, 2012).

In the latest models, PFA dynamics by complex are modelled using the estimates of adult spawners, adjusted to the number of eggs per fish based on life-history
characteristics of the age groups within each region of the stock complexes (ICES, 2011; Chaput, 2012). The spawner to PFA dynamic is modelled as:

$$
P F A_{y}=e^{\alpha_{y}} L E_{y} e^{\varepsilon}
$$

where: $\alpha_{y}$ is the productivity parameter from eggs ( $\times 1000$ ) to PFA (number of fish) for PFA year $y$ (on a log-scale), LEy the estimated lagged eggs ( $\times 1000$ ) corresponding to the PFA cohort in year $y$, and the progress of $\alpha_{y}$ is modelled as $a_{y+1}=a_{y}+\varepsilon$, with $\varepsilon \sim N\left(0, \sigma^{2}\right)$.

Productivity is modelled as an integration of survival in freshwater and during the first year at sea. An important assumption is the absence of heritability of age at maturity, i.e. all eggs are considered equivalent regardless of the age of the spawners. Lagged eggs refer to the adjustment of the egg depositions to correspond to the expected age at smoltification. Spawners in year ' $n$ ' contribute to recruitment in years ' $n+3$ ' to ' $n+8$ ' depending upon the relative proportions of one to six year-old smolts that they produce. For example, spawners in year ' $n$ ' produce eggs that hatch in year $n+1$ ' and may produce one year-old smolts in year ' $\mathrm{n}+2^{\prime}$ ', which would become 1 SW recruits in year ' $n+3$ '. Any two year-old smolts from the same spawners would produce 1 SW recruits in year ' $n+4$ ', etc.

At the stock complex level, lagged eggs are the sum of the eggs from the spawners in year $y-(s+2)$ weighted by the proportion of the smolts produced at age $s$ in region $k$ summed over regions in the complex. Two years are added to the smolt age, for the spawning year and smolt migration year to lag the eggs to the corresponding year of PFA:

### 3.3.2 NEAC PFA Forecast model

A forecast model to estimate PFA for all four NEAC stock complexes has been developed in a Bayesian framework by the Study Group on Salmon Stock Assessment and Forecasting (SGSSAFE). The model was originally reported to WGNAS in 2009 (ICES, 2009a), but was subsequently refined and has been in use by WGNAS in its present form since 2011 (ICES, 2011). The models for the northern and southern NEAC stock complexes have exactly the same structure and are run independently. A Directed Acyclical Graph (DAG) for the models is provided in Figure 3.3.2.1. The del considers both the maturing PFA (denoted PFAm) and the non-maturing PFA (denoted PFAnm). The full code used for running the model is available on the WGNAS SharePoint site.

A disaggregated version of the Bayesian model has since been developed using the same structure to provide forecasts at a country level, for all countries in both southern and northern NEAC model implementations (ICES, 2013). In these, countries are linked hierarchically only through the variance on the productivity parameter "a". There is no modelling linkage between the northern and southern complexes.


Figure 3.3.2.1. Directed Acyclical Graph (DAG) of the structure of the combined sea age model for the southern NEAC and northern NEAC forecast models. Ellipses in grey are observations (or pseudo-observations) derived from sampling programmes or from submodels (run-reconstruction).

The PFA is modelled using the summation of lagged eggs from 1SW and MSW fish $(L E)$ for each yeart $t$ and an exponential productivity parameter (a).


The productivity parameter (a) is the proportionality coefficient between lagged eggs and PFA. This is forecasted one year at a time $\left(a_{t-1}\right)$ in an auto correlated random walk, using the previous year's value (a) as the mean value in a normal distribution, with a common variance for the time-series of $a$.

$$
a t+1=a t+\varepsilon ; \quad \varepsilon \sim N(0, a . \sigma 2)
$$

The maturing PFA (denoted PFAm) and the non-maturing PFA (denoted PFAnm) recruitment streams are subsequently calculated from the proportion of $P F A$ maturing (p.PFAm) for each year t.p.PFAm is forecast as an autocorrelated value from a normal distribution based on a logit scale, using the previous year's value as the mean and a common variance across the time-series of p.PFAm.

```
logit.p.PFAmt+1 ~ N(logit.p.PFAmt , p.\sigma2)
logit.p.PFAmt = logit (p.PFAmt)
```

Uncertainties in the lagged eggs were accounted for by assuming that the lagged eggs of 1SW and MSW fish were normally distributed with means and standard deviations derived from the Monte-Carlo run-reconstruction at the scale of the stock complex. The uncertainties in the maturing and non-maturing PFA returns are derived in the Bayesian forecast models through the pseudo-observation method proposed by Michielsens et al. (2008), as used in the NAC model.

The natural mortality in the post-PFA time point was assumed constant among years, centred on an instantaneous rate value of 0.03 per month with a $95 \%$ confidence interval range of 0.02 to 0.04 .
Catches of salmon at sea in the West Greenland fishery (as 1SW non-maturing salmon) and at Faroes (as 1SW maturing and MSW salmon) were incorporated directly within the inference and forecast structure of the model, taking into account uncertainty in unreported catch at Faroes and the uncertainty in the allocation of catch at Greenland to NAC and NEAC based on sampling information. For southern NEAC, the data are available for a time-series of lagged eggs and returns commencing in 1978. Although the return estimates to southern NEAC begin in 1971, the lagged eggs are only available from 1978 due to the smolt age distributions (one to five years). For northern NEAC, data are available for a shorter time-series. Return and spawner estimates begin in 1983, but due to the smolt age distributions (one to six years), the lagged eggs are only available from 1991 onward. The models are fitted and forecasts derived in a consistent Bayesian framework.

The model provides forecasts for maturing and non-maturing stocks for both southern and northern NEAC complexes (and countries) for five years. Risks are defined each year as the posterior probability that the PFA would be greater than or equal to the age and stock complex/country specific Spawner Escapement Reserves (SERs), under the scenario of no exploitation

The country disaggregated version of the Bayesian NEAC inference and forecast model incorporates country specific catch proportions at Faroes, lagged eggs and returns of maturing and non-maturing components. Model structure and operation is as described above, incorporating country and year indexing. There is currently no hierarchical structuring of the countries within each stock complex. The evolution of a (the proportionality coefficient between lagged eggs and PFA) and its variance is adependent between countries. Similarly, the evolution of the proportion maturing PFAm) is also independent for each country, as is its variance.

### 3.3.3 NAC PFA Forecast model

WGNAS (ICES, 2009; 2012; 2015) developed forecasts of the pre-fishery abundance for the non-maturing 1SW salmon (PFA) using a Bayesian framework that incorporates the estimates of lagged spawners and works through the fisheries at sea to determine the corresponding returns of 2 SW salmon, conditioned by fisheries removals and natural mortality at sea. This model considers regionally-disaggregated lagged spawners and returns of 2SW salmon for the six regions of North America. The model is summarised in the Directed Acyclical Graph in Figure 3.3.1.1. The year is identified by the i index. The full code used for running the model is available on the WGNAS SharePoint site.


Figure 3.3.1.1. Directed Acyclical Graph (DAG) of the structure of the region disaggregated forecast model for 2SW salmon of North American origin. Ellipses in grey are observations (or pseudoobservations) derived from sampling programmes or from submodels (run-reconstruction).

Lagged spawners $L S_{i, k}$ represent the sum of smolt age adjusted annual spawners by region (k) that would be expected to contribute to the recruitment at sea prior to the fisheries (PFA) for year 1. LS ${ }_{i, k}$ are not directly observed but are estimated from the runreconstruction submodel used to estimates returns and spawners to each of the six regions.

The probability distributions of LS (and returns of $2 S W$ ) by region are used as likelinood functions expressing comparative degrees of belief given the data and a probability model not explicitly specified in the current model. The probability distributions were drawn from the Monte Carlo simulations and assumed to be normal with known mean (LS.m) and precision (1/variance) (tau.LS). The use of this distribution as a likelihood function is equivalent to assuming a pseudo-observation qual to LS.m issuing from a sampling distribution with mean and precision equal to LS and tau.LS (Michielsens et al., 2008).

$$
\text { LS. } \mathrm{m}_{\mathrm{i}, \mathrm{k}} \sim \mathrm{~N}\left(\mathrm{LS}_{\mathrm{i}, \mathrm{k}}, \text { tau. } \mathrm{LS}_{\mathrm{i}, \mathrm{k}}\right)
$$

The LS. $\mathrm{m}_{\mathrm{i}, \mathrm{k}}$ (mean) and tau. $\mathrm{LS}_{\mathrm{i}, \mathrm{k}}$ (precision) were derived assuming the lagged spawner values issued from a normal distribution characterized by the $95 \%$ confidence interval range statistics retained from the Monte Carlo simulations of returns.

Similarly, the returns of $2 S W$ salmon to the six regions ( $\mathrm{NR} 2_{\mathrm{i}, \mathrm{k}}$ ) are not directly observed, but estimated from the run-reconstruction model. The probability distributions were assumed to be normal with known mean NR2.m and variance tau.NR2. As with the LS variable, the NR2 were treated as pseudo-observations equal
to NR2.m issuing from normal sampling distributions with means and variances equal to NR2 and tau.NR2.

$$
\text { NR2. } \mathrm{m}_{\mathrm{i}, \mathrm{k}} \sim \mathrm{~N}\left(\mathrm{NR} 2_{\mathrm{i}, \mathrm{k},} \text { tau.NR2 } \mathrm{i}_{\mathrm{i}, \mathrm{k}}\right)
$$

In between the lagged spawners and returns as 2 SW salmon, the catches in the various sea fisheries and conditioning for natural mortality as the fish move from the time of the PFA to homewaters, are incorporated (Figure 3.3.1.1). The catches in the commercial fisheries of West Greenland and the Newfoundland and Labrador commercial and coastal fisheries (NG1.tot, NC1.tot and NC2.tot) are not directly observed, but estimated with error. The catches are converted to numbers of fish of 1SW non-maturing and 2SW fish based on the characteristics of the fish in the catch. Their (prior) probability distributions are obtained from catch statistics according to a formal structure included in the model.

Catches of large salmon (assumed to be 2SW salmon) from the St Pierre \& Miquelon fisheries (SPMC) are also included in the model as point estimates.

The natural mortality in the post-PFA time point was assumed constant between years, centred on an instantaneous rate value of 0.03 per month ( $95 \%$ confidence interval range of 0.02 to 0.04 )

For the NAC 2SW component, the modelis fitted to an historical dataseries of lagged eggs starting from 1978. Although the return and spawner estimates for NAC begin in 1971, the lagged eggs are only available from 1978 due to the smolt age distributions (one to six years).

The years are modelled independently conditional on the lagged spawners and yearly productivity parameters, The lagged spawners to PFA ratios (productivity) are modelled dynamically, i.e. assuming they are sequentially dependent within a region and attempts to take into account the most significant sources of uncertainty. The DAG for the model is shown in Figure 2.

PFA is assumed to be proportional to lagged-spawners (LS), with i.i.d. lognormal , and is modelled separately for each region $(k=6)$. The first year in the timeseries ( t ) is 1978 for lagged spawners (due to the range of smolt ages 1 to 6 for NAC and the start of the spawner time-series in 1970) and the last year of lagged spawner data is for the 2020 PFA year. The PFA can be modelled for 1978 to 2016 (the last PFA for which returns of 2SW salmon have been estimated back to rivers in 2017).

$$
\begin{aligned}
& P F A_{i, k}=\log N\left(\overline{P F A_{i, k}}, \sigma \cdot P F A^{2}\right) \\
& \overline{P F A_{i, k}}=\log \left(L S_{i, k}\right)+a_{i, k}+\varepsilon_{i, k} \\
& \text { iid } \\
& \varepsilon_{i, k} \sim N\left(0, \sigma \cdot P F A^{2}\right)
\end{aligned}
$$

The total PFA is calculated as the sum of the regional PFA's $(k=6)$. The proportion of the total PFA in each region is calculated directly as:
p. PFA $_{i, k}=$ PFA $_{i, k} /$ PFA.tot $_{i}$

A non-informative prior is assumed for $\sigma . P F A^{2}\left(1 / \sigma . P F A^{2} \sim \operatorname{gamma}(0.01,0.01)\right.$

The proportionality coefficient $(\log ) a_{i, k}$ between $\mathrm{LS}_{\mathrm{i}, \mathrm{k}}$ and $\mathrm{PFA}_{\mathrm{i}, \mathrm{k}}$ for each region is modelled dynamically as a random walk with a year and region residual variation ( $\eta_{i, k}$ ) assumed multivariate normal (MVN). The variance covariance matrix ( $\Sigma$ ) allows for correlations among regional productivity values reflecting that the fish share a common marine environment during part of their life cycle and that there are regional specificities in the evolution of the freshwater or the marine coastal environment.

$$
\begin{aligned}
a_{i+1, k} & =a_{i, k}+\eta_{i, k} \\
\eta_{i, k}^{i d} & \sim \operatorname{MVN}(0, \Sigma)
\end{aligned}
$$

The common yearly evolution of $a_{i}$ is the mean of annual a across regions:

$$
\mathrm{a} . \mathrm{y}_{\mathrm{i}}<-\operatorname{mean}\left(\mathrm{a}_{\mathrm{i}, \mathrm{k}}\right)
$$

This parameterization of the covariance of the proportionality coefficient differs from an earlier version of the model for which the proportionality (log) coefficient $a_{i, k}$ between $\mathrm{LS}_{\mathrm{i}, \mathrm{k}}$ and $\mathrm{PFA}_{\mathrm{i}, \mathrm{k}}$ for each region dynamically as a random walk model with the addition of a regionally common annually varying parameter (e.yi).

$$
\begin{aligned}
& a_{i+1, k}=a_{i, k}+e . y_{i+1}+\omega_{i, k} \\
& e . y_{i} \stackrel{i . i . d}{\sim} N\left(0, y \cdot \sigma^{2}\right)
\end{aligned}
$$

The correlation matrix of $a$ among the regions is calculated from the covariance matrix:

1) the precision matrix is inverted to produce the covariance matrix;
2) the covariance matrix is transformed to the correlation matrix.

The positive-definite matrix ( T , the precision matrix) is inverted:
variance matrix <- solve(T)
correlation matrix $<-\operatorname{cov} 2 \operatorname{cor}(b)$

The dynamic component of the model requires initialization for the first year ( $\mathrm{i}=1978$ )
-

$$
a_{1, k} \stackrel{i . i . d}{\sim} N(0,100)
$$

The models are fitted and forecasts derived in a consistent Bayesian framework under the OpenBUGS 3.0.3 software (http://mathstat.helsinki.fi/openbugs/; Lunn et al., 2000).

### 3.3.4 Summary of NAC and NEAC forecast models

The data inputs and models currently used by WGNAS for forecasting and provision of catch advice differ between the Commission areas; outline details are summarised in the text table below.

## FORECAST MODELS

|  | NAC | NEAC |
| :---: | :---: | :---: |
| Data inputs |  |  |
| Time period of data | 1978 on | 1978 on for southern NEAC |
|  |  | 1991 on for northern NEAC |
| Spatial aggregation | Separately for six regions of North America | By southern and northern stock complexes \& NEAC countries |
| Age components | 2SW salmon component only | 1SW and MSW age components |
| Spawners | Lagged spawners by region for 2SW salmon only | Lagged eggs by sea age component for the southern and northern complexes/country |
| Returns | Returns by region of $2 S W$ salmon only | Returns of 1SW and MSW age components by stock |
| Model structure |  |  |
| Spatial aggregation | Spawners and returns of 2SW salmon for six regions | Spawners and returns for two sea age components for the southern and northern <br> NEAC complexes/coumtries |
| Dynamic function | Random walk dynamic | Random walk dynamic |
|  | Region-specific recruitment rates linked with an annual recruitment rate variable | Sea age specific recruitment rates linked with a probability of maturing variable |
| Latent variables interest | PFA 1SW non-maturing <br> Recruitment rate by region and year | PFA 1SW maturing and PFA 1SW non-maturing by stock complex/ country <br> Recruitment rate by sea age component and the probability of maturing variable |
| Forecast years | our years | Five years; i.e. the present year -1 , the present year, and the next three years <br> ( $\mathrm{y}-1$ is a forecast, as the MSW stock component is yet to return). |

### 3.4 The development of a risk analysis framework for catch advice <br> 4.1 Introduction

provision of catch advice in a risk framework involves incorporating the uncertainty in all the factors used to develop the catch options (ICES, 2002). The ranges in the uncertainties of all the factors will result in assessments of differing levels of precision. The analysis of risk involves four steps:

1 ) identifying the sources of uncertainty;
2 ) describing the precision or imprecision of the assessment;
3 ) defining a management strategy; and
4 ) evaluating the probability of an event (either desirable or undesirable) resulting from the fishery action.

The uncertainties have been identified and quantified in the assessment of PFA for salmon stocks in both the NAC and NEAC areas. NASCO's strategy for the management of salmon fisheries is based upon the principle of ensuring that stocks are above CLs (defined in terms of spawner escapement or egg deposition) with a high probability. The undesirable event to be avoided is that the spawning escapement after the fisheries will be below the CLs.

### 3.4.2 Catch advice and risk analysis framework for the West Greenland fishery

A risk framework for the provision of catch advice for the West Greenland fishery has been applied since 2003 (ICES, 2003) and has been subject to a number of subsequent updates. The current procedure is outlined below. This involves estimating the uncertainty in meeting defined management objectives at different levels of catch (catch options). The risk framework has been formally accepted by NASCO.

Two stock complexes are of relevance to the management of the West Greenland fishery; non-maturing 1SW fish from North America and non-maturing 1SW fish from southern NEAC. The risk assessments for the two stoak complexes are developed in parallel and then combined at the end of the process into a single summary plot or catch options table. The risk analysis proceeds as illustrated in the flowchart in Figure 3.4.2.1).

The primary inputs to the risk analysis for the complex at West Greenland are:

- PFA forecast for the year of the fishery; PFANa and PFAneac;
- Harvest level being considered (t of salmon);
- Conservation spawning limits or alternate management objectives; and
- The post-fishery returns to each region.

The risk analysis of catch options incorporates the following input parameter uncertainties: (i) the uncertainty of the pre-fishery abundance forecast, (ii) the uncertainty in the biological parameters used to translate catches (weight) into numbers of salmon, and (iii) the uncertainty in attaining the conservation requirements simultaneously in different regions.

The uncertainty in the PFAna and PFAneac is accounted for in the forecast approach described above. The number of 1SW non-maturing fish of North American and European origin in a given catch $(\mathrm{t})$ is conditioned by the continent of origin of the fish (propna, prope), by the average weight of the fish in the fishery ( Wt tallages ), and by the proportion 1SW non-maturing fish in the respective continent of origin catches. These parameters define how many fish originating in North America and Europe are expected in the fishery harvests. For a level of fishery under consideration, the weight of the catch is converted to number of fish of each continent of origin using the following equation:

$$
\left.C 1 S W_{c}=\frac{t \text { X propC }}{A C F *(\text { propNAX Wt1SW }}+\operatorname{propE} X \text { Wt1SW }{ }_{E}\right) ~
$$

where: C1SWc is the catch (number of fish) of 1SW salmon originating in continent C (either North America or Europe), $t$ is the fishery harvest at West Greenland in kg ,
propC is the proportion of the 1 SW salmon harvest which originates from continent C , $W_{t 1 S W n A}$ and $W t 1 S W_{E}$ are the average weight $(\mathrm{kg})$ in the fishery of a 1 SW salmon of North American and European origin, respectively, and $A C F$ is the age correction factor by weight for salmon in the fishery which are not at age 1SW.

Since these parameters are not known for the forecast years of interest, they are estimated from previous values. Thus, propna (and propneac as 1 - propna) are drawn randomly from observed values of the past five years taking account of uncertainty due to sample sizes. For the other parameters, it is assumed that the parameters for $\mathrm{Wt}_{\text {Allages }}$ and the proportion non-maturing 1 SW in the catch by continent of origin could vary uniformly within the values observed in the past five years.

For a level of fishery under consideration, the weight of the catch is converted to fish of each continent of origin and subtracted from one of the simulated forecast values of PFAna and PFAneac. The fish that escape the Greenland fishery are immediately discounted by the fixed sharing fraction ( $\mathrm{F}_{\text {na }}$ ) historically used in the negotiations of the West Greenland fishery. The sharing fraction chosen is the 40\%:60\% West Greenland:North America split. The same sharing arrangement has been used for NEAC stocks (ICES, 2003). Any sharing fraction could be considered and incorporated at this stage of the risk assessment.

After the fishery, fish returning to homewaters are discounted for natural mortality from the time they leave West Greenland to the time they return to rivers. For North America this is a total of eleven months at a rate of $\mathrm{M}=0.03$ (equates to $28.1 \%$ mortality). For southern NEAC stocks this is a total of eight months at a rate of $\mathrm{M}=$ 0.03 (equates to $21 \%$ mortality). The fish that survive to North American homewaters are then distributed among the regions based on the regional proportions of lagged spawners for the last five years when estimates of spawners were available. The uncertainty in the regional proportions is characterised by drawing at random from a uniform distribution defined by the minimum and maximum regional ranges from the five years and calculating the average proportion for each of the six regions in North America.


Figure 3.4.2.1. Flow chart summarising risk analysis for catch options at West Greenland using the PFANA and the PFANEAC predictions for the year of the fishery. Inputs with solid borders are considered known without error. Estimated inputs with observation error that is incorporated in the analysis have dashed borders. Solid arrows are functions that introduce or transfer without error whereas dashed arrows transfer errors through the components.

The final step in the risk analysis of the catch options involves combining the conservation requirement or alternate management objectives with the probability distribution of the returns to North America for different catch options. Estimated 2SW returns to each region are compared to the conservation objectives of Labrador, Newfoundland, Québec, and Gulf. Estimated returns for Scotia-Fundy are compared to the objective of achieving an increase of $25 \%$ relative to average returns of the base period, 1992-1996. For the USA, the management objective was revised in 2014 (ICES, 2014). Estimated returns for the USA are now compared to the objective of achieving 2 SW adult returns of 4549 or greater. The advice to fisheries managers is presented as a probability plot (or table) of meeting or exceeding the objectives relative to increasing harvest levels at West Greenland.

ICES has adopted, a risk level of $75 \%$ of simultaneous attainment of seven management objectives (ICES, 2003) as part of an agreed management plan for the West Greenland fishery. The same level of risk aversion is applied for catch advice for homewater fisheries on the North American stock complex.

The catch advice for the West Greenland fishery is currently tabulated to show the probability of each management unit achieving its CL (or alternative reference level) individually and the probability of this being achieved by all management units simultaneously (i.e. in the same given year) (e.g. ICES, 2012a). This allows managers to evaluate both individual and simultaneous attainment levels in making their management decisions. Table 3.4.2.1 provides an example of catch options for West

Greenland for the years 2012 to 2014 (ICES, 2012a). An updated catch options table for 2015 to 2017 is provided in Section 5.3 of the 2015 Working Group report (see above).

The models currently used by WGNAS in developing catch advice are considered to provide consistent characterisation of the status and expectations for Atlantic salmon in the North Atlantic. Compared to previous models used by WGNAS prior to 2009, the Bayesian models provide more flexibility, are consistent with the emerging emphasis on such approaches in natural resource assessment, and can provide management advice consistent with the probability of achieving management objectives.
Table 3.4.2.1. Example of catch options tables for mixed-stock fishery at West Greenland by year of PFA, 2015 to 2017.

| 2015 | Probability of meeting or exceeding region-specific management objectives |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Catch option (t) | LAB | NFLD | QC | GULF | SF | USA | S-NEA | ALL |
| 0 | 0.45 | 0.86 | 0.71 | 0.50 |  | 0.89 | 0.9 | 0.06 |
| 10 | 0.42 | 0.84 | 0.67 | 0.48 | 0.14 | 0.88 | 0.98 | 0.05 |
| 20 | 0.40 | 0.83 | 0.63 | 0.45 | 0.13 | 0.87 | 0.98 | 0.05 |
| 30 | 0.38 | 0.81 | 0.59 | 0.42 | 12 | 0.8 | 0.98 | 0.04 |
| 40 | 0.36 | 0.78 | 0.54 |  | 0.12 | 0.83 | 0.98 | 0.04 |
| 50 | 0.34 | 0.76 | 0.50 | 0.38 | 0.11 | 0.81 | 0.98 | 0.03 |
| 60 | 0.32 | 0.73 | 0.46 |  | 0.10 | 0.79 | 0.98 | 0.03 |
| 70 | 0.30 | 0.70 | 0.42 | 0.3 | 0.09 | 0.77 | 0.98 | 0.03 |
| 80 | 0.28 | 0.67 | 0.39 |  | 0.08 | 0.74 | 0.98 | 0.03 |
| 90 | 0.26 | 64 |  | 0.29 | 0.08 | 0.72 | 0.97 | 0.02 |
| 100 |  | 0.60 |  | 0.27 | 0.07 | 0.68 | 0.97 | 0.02 |
| 2016 Catch | Prob | ity of m | or | eding reg | spe | man | ment objec |  |
| Option (t) | LAB | FFL | QC | GULF | SF | USA | S-NEAC | ALL |
| 0 |  | 0.78 | 0.73 | 0.50 | 0.25 | 0.75 | 0.95 | 0.08 |
| 10 | 0.46 | 0.76 | 0.70 | 0.48 | 0.24 | 0.73 | 0.95 | 0.07 |
|  | 0.44 | 0.75 | 0.67 | 0.46 | 0.23 | 0.72 | 0.95 | 0.06 |
| 30 |  | 0.73 | 0.63 | 0.44 | 0.22 | 0.70 | 0.95 | 0.06 |
|  | 0.41 | 0.70 | 0.60 | 0.42 | 0.21 | 0.68 | 0.95 | 0.06 |
| 50 | 0.39 | 0.68 | 0.56 | 0.40 | 0.20 | 0.66 | 0.94 | 0.05 |
|  | 0.37 | 0.65 | 0.53 | 0.38 | 0.19 | 0.64 | 0.94 | 0.05 |
| 70 | 0.35 | 0.63 | 0.50 | 0.36 | 0.18 | 0.62 | 0.94 | 0.05 |
| 80 | 0.33 | 0.60 | 0.47 | 0.34 | 0.17 | 0.59 | 0.94 | 0.04 |
| 90 | 0.31 | 0.57 | 0.44 | 0.32 | 0.16 | 0.57 | 0.94 | 0.04 |
| 100 | 0.30 | 0.54 | 0.41 | 0.31 | 0.15 | 0.55 | 0.94 | 0.04 |
| 20170 | Probability of meeting or exceeding region-specific management objectives |  |  |  |  |  |  |  |
| Catch <br> Option ( t ) | LAB | NFLD | QC | GULF | SF | USA | S-NEAC | ALL |
| 0 | 0.56 | 0.78 | 0.75 | 0.55 | 0.20 | 0.86 | 0.94 | 0.08 |
| 10 | 0.55 | 0.77 | 0.73 | 0.53 | 0.20 | 0.85 | 0.94 | 0.08 |
| 20 | 0.53 | 0.75 | 0.70 | 0.51 | 0.19 | 0.84 | 0.94 | 0.07 |
| 30 | 0.52 | 0.73 | 0.67 | 0.49 | 0.18 | 0.83 | 0.94 | 0.07 |


| 2015 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Catch <br> option (t) | Probability of meeting or exceeding region-specific management objectives |  |  |  |  |  |  |  |
|  | LAB | NFLD | QC | GULF | SF | USA | S-NEAC | ALL |
| 40 | 0.50 | 0.71 | 0.64 | 0.47 | 0.17 | 0.82 | 0.94 | 0.06 |
| 50 | 0.48 | 0.69 | 0.62 | 0.46 | 0.17 | 0.81 | 0.94 | 0.06 |
| 60 | 0.46 | 0.67 | 0.59 | 0.44 | 0.16 | 0.79 | 0.94 | 0.06 |
| 70 | 0.45 | 0.65 | 0.56 | 0.42 | 0.16 | 0.77 | 0.94 | 0.05 |
| 80 | 0.43 | 0.63 | 0.54 | 0.41 | 0.15 | 0.76 | 0.94 | 0.05 |
| 90 | 0.42 | 0.61 | 0.51 | 0.39 | 0.14 | 0.74 | 0.94 | 0.05 |
| 100 | 0.40 | 0.59 | 0.49 | 0.38 | 0.14 | 0.72 | 0.94 | 0.05 |

### 3.4.3 Catch advice and risk analysis framework for the Faroes fishery

### 3.4.3.1 Outline of the risk framework and management decisions required

There is currently no agreed framework for the provision of catch advice for the Faroes fishery adopted by NASCO. However, NASCO has asked ICES, for a number of years, to provide catch options or alternative management advice with an assessment of risks relative to the objective of exceeding stock conservation limits for salmon in the NEAC area. An initial risk framework that could be used to provide and evaluate catch options for the Faroes fishery was outlined by WGNAS in 2010 (ICES, 2010a). This was based on the method currently used to provide catch advice for the West Greenland fishery, which involves estimating the uncertainty in meeting defined management objectives at different catch levels (TAC options). The Faroes risk framework was developed further at subsequent WGNAS meetings (ICES, 2011; 2012a) and the current proposed procedure is outlined below.

A number of decisions are required by managers before full catch advice could be provided (ICES, 2011; 2012a). Specifically, ICES has indicated that NASCO would need to agree upon the following issues before the risk framework could be finalised:

- season (January to December or October to May) to which any TAC should apply
share arrangement for the Faroes fishery;
choice of management units for NEAC stocks; and
specification of management objectives.

In developing an indicative risk framework, WGNAS has made pragmatic choices regarding these issues:

Faroes fishing season: A decision is required on the period to which any TAC for the Faroes fishery would apply. The Faroes fishery has historically operated between October/November and May/June, but the historical TACs applied to a calendar year. This means that two different cohorts of salmon of each age class (e.g. two cohorts of 1SW salmon, etc.) were exploited under each TAC. ICES (2011) recommended that NASCO manage any fishery on the basis of fishing seasons operating from October to June, and catch advice should be provided on this basis.

Sharing agreement: The 'sharing agreement' establishes the proportion of any harvestable surplus within the NEAC area that could be made available to the Faroes
fishery through the TAC. In effect, for any TAC option being evaluated for the Faroes, it is assumed that the total harvest would be the TAC divided by the Faroes share. WGNAS has proposed using a share allocation derived using the same approach and baseline period (1986-1990) as for West Greenland (ICES, 2010a). This gave a potential share allocation of $7.5 \%$ to Faroes. Following discussion within NASCO, one Party proposed an alternative baseline period of 1984-1988, which would give a share allocation of $8.4 \%$ to Faroes. In the absence of further advice from NASCO, WGNAS has applied a value of $8.4 \%$.

Choice of management units: ICES (2010a) noted that the stock complexes currently used for the provision of NEAC catch advice (southern NEAC and northern NEAC) are significantly larger than each of the six management units used for North American salmon (2SW only) in the catch advice for the West Greenland fishery. Basing an assessment of stock status on these large units greatly increases the risks to indivjdual NEAC river stocks or groups of stocks that are already in a more depleted state than the average.

For the provision of catch advice on the West Greenland fishery, the total CL for NAC (2SW salmon only) of about 152000 fish is assessed in six management units, which means that each unit has an average CL of about 25000 salmon. In contrast, the total CLs for each of the NEAC stock complexes are:


The NEAC stock complexes are therefore between eight and 25 times the size of the average NAC ones. There is also wide variation in the size and status of stocks both within and among the NEAC national stock groups. WGNAS recommended (ICES, 2012a) that the NEAC catch advice should be based on more management units than but also noted that there are practical limitations on the extent to which the assessments can be disaggregated, since the availability of information on the composition of the catch at Faroes constrained the selection of management units. In 2013, WGNAS (ICES, 2013) proposed a method to estimate the stock composition of the Faroes catch at a national level based on tag returns and the PFA estimates, but did not consider it appropriate to extend this to stock complexes smaller than this. Genetic stock assignment studies are underway to analyse scale samples collected at Faroes, but these are not expected to facilitate disaggregation below this level. In addition, other parameter values used in the assessment are currently only available for the total fishery and not smaller stock complexes.

In providing indicative catch advice with the new framework, WGNAS considered that it would be informative to managers to provide catch options tables for the ten NEAC countries as well as for the four stock complexes and has therefore run the risk framework using management units based on countries.

Management objectives: The management objectives provide the basis for determining the risks to stocks in each management unit associated with different catch options. The NASCO agreement on the adoption of a Precautionary Approach (NASCO, 1998) calls for the 'formulation of pre-agreed management actions in the form of procedures to be applied over a range of stock conditions', indicating that the management objectives (e.g. the required probability of exceeding the CL) should be agreed in advance of specific management proposals being considered.

At the request of NASCO, WGNAS considered the implications of applying probabilities of achieving CLs to separate management units vs. the use of simultaneous probabilities; this issue was outlined in detail in ICES (2013)

The probability of simultaneous attainment of management objectives in a number of separate management units is roughly equal to the product of the probabilities of individual attainment for each management unit. The probability threshold for each individual management unit might reasonably be set at a fixed level unless there are specific reasons for adopting an alternative (e.g. for stock rebuilding) ICES (2012) recommended that an appropriate probability level for individual stock complexes would be $95 \%$. This individual probability level can be applied to each management unit regardless of the number of units used; however, this is less obvious for the probability of simultaneous attainment, as explained next.
Management decisions for the West Creenland fishery have been based on a $75 \%$ probability of simultaneous attainment of LS. For a given probability of achieving individual stock CLs, the probability of simultaneous attainment decreases rapidly as the number of management units considered increases. For the example of 20 management units (e.g. two age groups from each of ten countries), the use of the simultaneous probability level applied for West Greenland (75\%) would correspond to the probability of individual stocks meeting the CLs being $98.6 \%$ or higher, assuming the same individual probability for all stocks. The use of a $95 \%$ probability level for meeting the CLs individually in the 20 management unit example, implies a simultaneous attainment probability of about $36 \%$, i.e. there would be a $64 \%$ chance that at least one stock failed to meet its CL in any given year. On the other hand, the a $\mathbf{7 5 \%}$ probability of simultaneous attainment could result in a fishery being advised when the individual probability of one management unit is as low as $75 \%$ if all the othermanagement units have a $100 \%$ chance of meeting the CL (as in that case, the probability of simultaneous attainment would still be $75 \%$ ). This may not be an acceptable risk for managing multiple river stocks.

WGNAS considered that the probability of simultaneous attainment can provide useful information to managers of the risk of failing to meet CLs in at least one stock in the MSF. However, as the management units being considered by NASCO for managing the MSF at Faroes are still very large and each unit encompasses a large number of individual river stocks, choosing a high probability level (such as $95 \%$ ) of attaining CLs in individual units would be less risky to individual stocks than the use of a simultaneous attainment objective set at the value used for the West Greenland fishery.

On the basis of these considerations, WGNAS provided both individual probabilities and the probability of simultaneous attainment of the management units in the catch
options tables (ICES, 2013). ICES recommends that management decisions should be based principally on a $95 \%$ probability of attainment of CLs in each stock complex individually. The simultaneous attainment probability may also be used as a guide, but managers should be aware that this will generally be quite low when large numbers of management units are used (as illustrated above, in the example with 20 management units).

### 3.4.3.2 Modelling approach for the catch options risk framework

The basic model for assessing each catch option within the risk framework is the same for both stock complexes and at a country level (ICES, 2013). The PFA forecasts derived in the Winbugs model are transferred to the risk framework model run in ' R '. The estimates and distributions of the PFA estimates used in the risk framework are derived by taking the first 50000 values from the Winbugs posterior forecast simulations. Parameters in the following description that are marked with an" in the equations have uncertainty around them generated by means of 50000 random draws from the annual values observed from the sampling programmes conducted in the Faroes between the 1983/1984 and 1990/1991 fishing seasons. They therefore contribute to the estimation of the probability density function around the potential total harvest arising from each TAC option. When the assessment is run at a national level, the number of draws has to be limited to 25000 because of memory limitations in ' $R$ '.

The modelling procedure involves:

- estimating the total number of 1 SW and MSW salmon that could be killed as a result of any TAC at Faroes, including catches in homewaters;
- adjusting these to their equivalent nambers at the time of recruitment to the Faroes fish
- subtracting these from the PFA estimates for maturing and non-maturing 1SW salmon in the appropriate years;
- assessing the results against the SERs (i.e. the CLs adjusted to the time of recruitment to the Faroes fishery).

TACoption (T) is first divided by the mean weight $\left(\mathrm{Wt}^{*}\right)$ of salmon caught in the fishery to give the number of fish that would be caught, and this value is converted to numbers of wild fish $(\mathrm{Nw})$ by multiplying by one minus the proportion fish-farm escapees in samples taken from the Faroes catch ( $\mathrm{pE}^{*}$ ) observed in historical sampling programmes. A correction factor ( $C=0.63$ ) is applied to the proportion of fish-farm escapees to take account of reductions in the numbers of farm escapees over the past 20 years based on observations in Norwegian coastal waters:

$$
\mathrm{Nw}=\mathrm{T} / \mathrm{Wt}^{*} \mathrm{x}\left(1-\left(\mathrm{pE}^{*} \times \mathrm{C}\right)\right)
$$

This value is split into numbers by sea age classes (1SW and MSW) according to the proportion of each age group ( $\mathrm{pAi}^{*}$, where ' i ' is 1 SW or MSW) observed in historical catch sampling programmes at Faroes. In the past, there has also been a requirement to discard any fish less than 60 cm total length caught in the Faroes fishery, and $80 \%$ of these fish were estimated to die, so these mortalities are also added to the 1SW catch. Thus:

Nw1SW $=$ Nwtotal $\times \mathrm{pA}^{2} \mathrm{NW}^{*}+\left(\right.$ Nwtotal $\left.\times \mathrm{pD}^{*} /\left(1-\mathrm{pD}^{*}\right) \times 0.8\right)$
and
NwMSW = Nwtotal $\times$ pAMSW*
where: $\mathrm{pD}^{*}$ is the proportion of the total catch that is discarded (i.e. fish $<60 \mathrm{~cm}$ total length).

Further corrections are made to the 1SW and MSW numbers to reduce the 1SW total to take account of the proportion that will not mature as 1SW fish and to add the survivors from this group to the MSW fish in the following year. For the first catch advice year the number added to the MSW total is adjusted to the TAC of the current season (i.e. zero in 2012/2013). Thus:

$$
\begin{aligned}
& \mathrm{Nw} 1 \mathrm{SW}=\mathrm{Nw} 1 \mathrm{SW} \times \mathrm{pK} \text { * } \\
& \text { and }
\end{aligned}
$$

NwMSW = NwMSW + Nw1SW x (1-pK*)
where ' $p K$ ' is the proportion of 1 SW salmon that are expected to mature in the same year (0.78) derived from experimental studies conducted in the 1980s (Youngson and Webb, 1993).

The numbers in each age group are then divided among the management units by multiplying by the appropriate proportions ( pUij ), where ' i ' denotes the age groups and ' $j$ ' denotes the management units, and each of these values is raised by the Faroes share allocation (S) to give the total potential harvest (Hij) of fish from each management unit and sea age group:
$\mathrm{Nwij}=(\mathrm{Nwi} \times \mathrm{pUij}) / S$

Finally, these values are adjusted for natural mortality so that they can be compared with the PFA forecasts and SER values from the mid-date of the fishery to the recruitment date by using an instantaneous monthly rate of mortality of 0.03.

These harvests are then subtracted from the stock forecasts (PFAij) for the management units and sea age groups and compared with the Spawner Escapement Reserves (SERij) to evaluate attainment of the management objective. In practice, the attainment of the management objective is assessed by determining the probability that PFAij - Hij SERij is greater than zero. The SER is the number of fish that need to be alive at the time of the Faroes fishery to meet the CL when the fish return to homewaters; this equals the CL raised by the mortality over the intervening time. CLs and SERs are currently estimated without uncertainty.

### 3.4.3.3 Input data for the risk framework

The analysis estimates probability of each management unit achieving its SER (the overall abundance objective) for different catch options in the Faroes fishery (from 0 to 200 t ). The analysis assumes:

- no fishery operated for the most recent season;
- the TAC allocated to Faroes is the same in each year and is taken in full;
- homewater fisheries also take their catch allocation in full.

The analysis requires the following input data for the catch that would occur at the Faroes if a TAC was allocated (full details are provided in ICES, 2013):

- mean weights;
- proportion by sea age;
- discard rates (fish less than 60 cm total length);
- proportion of fish-farm escapees;
- composition of catches by management unit;
- proportion of 1SW fish not maturing.


### 3.4.3.4 Indicative catch advice

Table 3.4.3.4.1 provides an example of catch options for the Farges fishery for the seasons 2013/2014 to 2015/2016 (ICES, 2013). Equivalent tables were provided for both 1SW and MSW salmon for all NEAC countries, and WGNAS also estimates the exploitation rates that these TAC options would impose on each stock complex or national stock (ICES, 2013). Updated catch options tables forthe seasons 2015/2016 to 2017/2018 are provided in Section 3.5 of the 2015 Working Group report (see above).

Table 3.4.3.4.1. Probability of northern and southern NEAC-1SW and MSW stock complexes achieving their SERs independently and simultaneously for different catch options for the Faroes fishery in the 2013/2014 to 2015/2016 fishing seasons


### 3.5 Development of indicator frameworks to identify significant changes in previously provided multiannual management advice

### 3.5.1 Background

In support of the multiannual management advice that is provided for all three NASCO Commission Areas, NASCO asked ICES to provide an assessment of the minimal information needed to signal an unforeseen change in productivity for stocks contributing to fisheries within each Commission area. A particular concern was that an increase in productivity may alter the reliability of the previously provided multiyear catch options and could result in unrealised harvest within various mixed-
stock fisheries. Initial progress on this issue was presented to WGNAS in 2006 (ICES, 2006) and further developments were made by the Study Group on Establishing a Framework of Indicators of Salmon Stock Abundance [SGEFISSA] which met in November 2006 (ICES, 2007b) and reported to WGNAS in 2007 (ICES, 2007a). This resulted in the development of a suggested framework (Framework of Indicators FWI) which could be used to indicate if any significant change in the status of stocks had occurred and thus confirm whether the previously provided multi-annual management advice was still appropriate.

The initial FWI was developed with both the Greenland and Faroes fisheries in mind, although the methodology only proved suitable for the West Greenland fishery and an alternative approach was subsequently developed for the NEAC area (ICES, 2011; 2012a; 2013). Thus, FWIs are now routinely applied in the interim (non-assessment) years of multiyear agreements for both NAC and NEAC to facilitate the management of the West Greenland and Faroes fisheries respectively. Both operate according to the timeline outlined in Figure 3.5.1.1. Outline descriptions of the two different schemes are provided below.


Figure 3.5.1.1. Timeline for employment of the Framework of Indicators (FWI). In Year i, ICES provides an updated FWI which re-evaluates the updated datasets and is summarized in an Excel worksheet. In January of Year $\mathbf{i}+1$ the FWI is applied and two options are available depending on results. If no significant change is detected, no re-assessment is necessary and the cycle continues to Year $i+2$. If no significant change is detected in Year $i+2$, the cycle continues to Year $i+3$. If a significant change is detected in any year, then reassessment is recommended. In that case, ICES would provide an updated FWI the following May. ICES would also provide an updated FWI if year equals 4. [MACO = multi-annual catch options].

### 3.5.2 Framework of Indicators (FWI) for the West Greenland Fishery

The process for developing and applying the FWI for the Greenland fishery consists of six general steps:

- Definition of a significant change - Define measurable criteria for what the statement "a significant change in the previously provided multi-annual management advice" represents.
- Evaluating historical relationships between indicators and variables of interest - Define and evaluate the historical relationships between numerous indicators and the variable of interest for individual rivers across all stock complexes.
- Establishing threshold values - Define the threshold level (i.e. variable of interest level) that will satisfy the management objectives for each stock complex.
- Decision rule determinations - Define and apply a standardised approach for determining the appropriate decision rule value. The decision rule should provide a signal if the variable of interest will be greater than or less than the threshold level with high precision.
- Combining Indicators within the Framework - Define and apply standardised approach for combining indicator datasets within and acro stock complexes for future comparison against contemporary indicator values.
- Applying the FWI - Define and apply a standard approach to input contemporary indicator values into the FWI to determine if there is likely to be a significant change in the previously provided management advice.

Each of these is considered in turn; full details are available in ICES (2007b).

### 3.5.2.1 Definition of a significant change

A significant change in the previously provided multiannual management advice is regarded as an unforeseen change in stock status that would alter the previously provided advice based on analysis of current population data obtained from various monitored populations across the North Atlantic. This would be indicated by a situation where stock abundance has increased to a level where a fishery could be recommended when no eatch had previously been advised, or a decrease in stock abundance when catchoptions had been chosen.

For the fishery at West Greenland, ICES would recommend that a harvestable surplus exists within the West Greenland stock complex if there was a high probability ( $75 \%$ ) that the following three objectives could be met simultaneously:

The conservation limits of the four northern regions of North America (Labrador, Newfoundland, Québec, and Gulf) were achieved.

There was a $25 \%$ increase in returns to the Scotia-Fundy region relative to the mean returns for the 1992-1996 period. For the USA, the management objective was revised in 2014 to correspond to recover objectives defined in the recovery plan for endangered Atlantic salmon stocks in the USA (ICES, 2014), this now requires that estimated 2SW adult returns are 4549 or greater.

The conservation limit for the Southern NEAC MSW complex was achieved.

### 3.5.2.2 Evaluating historical relationships between indicators and variable of interest

A number of variables were considered for inclusion as indicators in the FWI, but only two were considered sufficiently informative to be carried forward into the framework: adult returns (returns, catch or estimated PFA) and return rates (i.e. smolt survival
rates, marine survival). These are available, by sea age class, for a number of monitored rivers throughout the North Atlantic and can be directly related to the management objectives for the fishery.

### 3.5.2.3 Establishing threshold values

In keeping with the $75 \%$ probability of meeting or exceeding the objectives for the West Greenland catch options (see above), the 25th percentile of the return estimates of the six areas in North America are compared to the corresponding 2SW conservation limits of the four northern areas of North America, to the $25 \%$ increase objective for the ScotiaFundy area, and to management objective of achieving 4549 or greater 2SW adult returns for the USA. For the southern NEAC non-maturing component, the 25th percentile of the PFA estimate of the southern NEAC non-maturing complex is compared to the spawning escapement reserve (SER) for the southern NEAC nonmaturing complex.

### 3.5.2.4 Decision rule determinations

The procedure for analysing the relationships between the indicators and the returns of 2SW salmon or the PFA estimates was originally suggested by ICES (2006). The individual river catches, returns or return rates are lagged to correspond to the same smolt cohort for the 2SW returns to North America or to the PFA estimates for NEAC complexes. Bivariate plots of each indicator dataset relative to the 2SW returns or the PFA estimates are prepared. Upper and lower halves are defined by the management objective value for the corresponding geographic area in North America or the NEAC stock complexes as outlined above. Estimates of returns of 2SW or PFA estimates in the upper half correspond to years when the returns or PFA exceed the management objectives. Points in the lower half correspond to years when the returns or PFA are less than the management objectiv

Left and right halves are defined by a sliding rule along the indicator range. An objective function that maximises the number of correct assignments (true highs and true lows) is used to define the indicator decision rule. The objective function also minimises the number of incorrect assignments (false highs and false lows, Figure



Figure 3.5.2.4.1. Example of Indicator/Variable of Interest exploratory graph identifying the threshold value, decision rule, penalty function and the four states (true high, true low, false high and false low).

The value of the indicator variable that minimises the sum of the penalty scores (i.e. minimises the number of incorrect assignments) is assigned as the decision rule for that dataset. Equal penalty weights are assigned to false highs (lower right quadrant) and false lows (upper left quadrant). Correet assignments are scored as zero. In the case when multiple minima occurred, the lowest indicator value among the low minima values is chosen.

Indicators are retained in the framework when they are evaluated as being informative of the magnitude of returns or PFA relative to the management objectives. These informative indicator datasets also have to meet the following two criteria to be retained:

Expectation that the indicator variable would be available in future (in January), and
A minimum of five observations are present in each of the correct quadrats (true low; true high).

### 3.5.2.5 Combining Indicators within the Framework

The probabilities of correct assignments are calculated for each of the true low and true high states for each of the indicator datasets retained. The respective probabilities correspond to the ratio of the correct assignment to all observations within the respective low and high indicator halves:
$\mathrm{P}($ Statelow I Indicatorlow) (i.e. true low) $=\mathrm{N}($ Statelow I Indicatorlow) / N Indicatorlow

```
P(Statehigh | Indicatorhigh) (i.e. true high) = N(Statehigh | Indicatorhigh) / N Indicatorhigh
```

Indicator datasets are then pooled according to management objective/stock complex groupings. Each NAC stock complex ( $\mathrm{n}=6$ ) and the southern NEC non-maturing stock complex are pooled separately as these stock complexes relate to the management objectives for the West Greenland fishery.

### 3.5.2.6 Applying the FWI

To apply the FWI, the most recent year's indicator value for each of the retained indicator datasets is compared to the decision rule as determined from the historical datasets. If the contemporary indicator value is low relative to the decision rule, it is assigned a value of -1 . If the value is high, it is assigned a score of +1 . Multiple indicators within the stock complex groupings are then combined by arithmetic average of the product of the indicator value $(-1,+1)$ and the probability of a correct assignment corresponding to the true low or true high states. An average geographic area or stock complex score equal to or greater than zeno suggests there is a likelihood of meeting the management objective for that grouping based on the historic relation between the variable of interest (adults return or PFA) and the indicators evaluated.

If the scores for all the groupings within a fishery complex are greater than zero, then there is a likelihood that all the management objectives for that fishery will be met. Under that scenario, the multiyear management advice should be reassessed. When the score(s) for one or more of the groupings is less than zero, there is unlikely to be a significant change in the management advice and there would be no need for a reassessment.

SGEFISSA (ICES, 2007b) developed a spreadsheet template FWI (see example at Figure 3.5.2.6.1) in which the underlying variable of interest/ indicator dataset relationships and decisionrules are summarised and collated according to the specific management objectives for each fishery. This provides one of two conclusions for the user:

1) No significant change identified by the indicators;


Figure 3.5.2.6.1. Framework of indicators spreadsheet for the West Greenland fishery, updated in April 2018 and applied in January 2019.

If no significant change has been identified by the indicators, then the multiyear catch advice for the year of interest could be retained. If a significant change is signalled by indicators, the response is to reassess.

The framework spreadsheet is designed to capture both fishing and non-fishing scenarios:

- Multiyear advice provides no catch options greater than zero but indicators are suggesting that the management objectives may be met (conclusion: Reassess);
- Multiyear advice provides catch options greater than zero but the indicators suggest the management objectives may not be met (conclusion: Reassess).
There are two steps required by the user to run the framework. The first step in the framework evaluation is to enter the catch advice option (i.e. tonnes of catch) for the

West Greenland fishery. This feature provides the two way evaluation of whether a change in management advice may be expected and a reassessment would be required. The second step is to enter the values for the indicator variables in the framework for the year of interest. The spreadsheet evaluation update is automated and the conclusion is shown in the row underneath "Overall Recommendation".

The conclusions from the framework evaluation are based on whether there is simultaneous achievement of the management objectives in the six stock areas of North America and the southern NEAC 1SW non-maturing complex. If there are no indicator variables for a geographic area, the attainment of the management objectives is evaluated as unknown and that area or complex is not used in the deeision structure of the framework.

Within the geographic areas for which indicator variables are retained, all the available indicators are used to assess the indicator score. If an update value for an indicator variable is not available for the year of interest, the indicator variable is not used to quantify the indicator score for that area.

The West Greenland FWI was updated during the 2018 WGNAS meeting and an updated spreadsheet produced. Details are provided in Section 5.8 of the Working Group report (see above).

### 3.5.3 Framework of Indicators (FWI) for the Faroes Fishery

### 3.5.3.1 Background

The original FWI applied to the West Greenland fishery (ICES, 2007b) was not applicable for the Faroese fishery for a number of different reasons. Among these were the lack of quantitative catch advice, the absence of specific management objectives and a sharing agreement for this fishery and the fact that none of the available indicator datasets met the criteria for inclusion in the FWI.

In 2011. WGNAS re-evaluated the approach for developing a FWI for the Faroese fishery (ICES, 2011). Since over the available time-series the PFA estimates for the NEAC stock complexes have predominately remained above the SER, the Working Group suggested a different set of decision rules for this FWI. It was suggested that the status of stocks should be re-evaluated if the FWI suggests that the PFA estimates are leviating substantially from the median values from the forecast.

### 3.5.3.2 Description of the FWI

It was initially suggested that the $95 \%$ confidence interval range for the mean of the indicator prediction, relative to the median forecast value, be used to compute the decision thresholds for whether the indicator suggests a reassessment or not (ICES, 2011). The limits should be computed at the median values of the PFA forecasts in each of the years in multiyear advice. However, the $95 \%$ criterion was subsequently reexamined (ICES, 2012a) and it was recommended that the upper and lower $75 \%$ confidence limits of the individual predictions be used for comparison (Figure 3.5.3.2.1). WGNAS recognised that this was a relaxation of the decision rule suggested in 2011, and will lead to a larger interval, and thus a lower chance of a reassessment than the approach suggested in 2011. However, this was considered to be a more
realistic confidence level given the relatively wide variability of the indicator datasets, and was also consistent with the approach adopted by NAC.


Figure 3.5.3.2.1. Example of how the reassessment intervals for the indicators are computed. The values of an indicator (counts) are plotted against the PFA. The regression line is shown in black and $75 \%$ confidence limits for the individual estimates are shown in red. From the forecasted PFA in the year in question the values of the indicator corresponding to the upper and lower $75 \%$ confidence interval are estimated. If the indicator value falls outside these limits a reassessment is recommended by this particular indicator.

When the stocks are divided into smaller management units, potential indicators for each management unit become relatively scarce. Therefore, the Working Group recommended that the FWI be regressed against the stock complexes that they belong to. For example, MSW indicators from Norway should be regressed against PFA MSW for northern NEAC.

2012, the FWI was applied as a two-tailed test (ICES, 2012a). However, it was subsequently agreed that, in the event of a closed fishery, the indicators should only be compared to the upper $75 \%$ confidence limit (i.e. a one-tailed test). This means that for a closed fishery, a reassessment is only triggered where the forecast appeared to be an underestimate and there may be a possibility of a harvest being denied. In the case of an open fishery they should be compared to both the upper and lower $75 \%$ confidence limits. In this case, if the FWIs suggest that the forecasted PFA is either an underestimation or an overestimation of the realised PFA in any of the four stock complexes, then this should trigger a reassessment.

ICES further advised (ICES, 2015) that, in the case of closed fisheries, the FWI should only be applied to those stock complexes which had previously signalled zero catch options at Faroes. This was agreed by NASCO (NEA(16)11) and applied in January 2017.

WGNAS developed a FWI spreadsheet (ICES, 2011) to provide an automatic evaluation of the need for a reassessment once the new indicator values are available in January; this has been updated in subsequent years (ICES, 2012a; 2013, 2015). An example spreadsheet is provided at Figure 3.5.3.2.2.

The following summarizes the main steps performed by the spreadsheet following updating of the relevant data for the variable of interest by adding the latest year's number:

- Regression analysis with the dataset $x$ to determine its power to predict PFA in the forecasted years.
- Calculation of the $75 \%$ confidence intervals of individual predictions of the regression for dataset $x$. An indicator value below the $75 \%$ individual confidence interval (CI) is interpreted as indicative of an overestimation of th PFA, while a point above the $75 \%$ individual confidence intervalis interpreted as indicative of an underestimation of PFA.
- A dataset is considered informative and should be kept as an indicator in the FWI if the following conditions are met: sample size $(n) \geq 10 ; r^{2} \geq 0.2$; dataset updated annually and new value available by January 15. Datasets that do not meet these criteria are discarded.
- Apply a binary score to each indicator value. Thus, for dataset $x$, if the current year's indicator value is outside the $75 \%$ individual regression point estimate Cl (below or above) then that indicator receives a seore of 1 . If the indicator is within the $75 \% \mathrm{Cl}$, then the indicator receives a score of -1 . In the absence of an indicator datapoint for any year, a score of zero is applied. Whether the indicator value is above or below the upper and lower Cl values is checked separately in two spreadsheet columns and a decision whether the indicator value is within the Cl is assessed by combining the information in the two columns
- Separate columns are used to sum the scores for all the indicator datasets within each stock complex. This is done separately for points that fall above he Cl and those that fall below. In the case of a two-sided approach (open fishery), if the sum of these columns is $\geq 0$, then the spreadsheet signals "REASSESS", if the sum is $<0$, then it signals "No significant OVERestimation of PFA identified by indicators, do not reassess" for indicator values that fall below the Cl , and "No significant UNDERestimation of PFA identified by indicators, do not reassess" for indicator values that are above the CI . In case of a one-sided approach (closed fishery), only underestimation will signal a "REASSESS".
- FWI results are generated for each stock complex (northern NEAC maturing and non-maturing, and southern NEAC maturing and non-maturing). A score of $\geq 0$ for any of these stock complexes would signal a reassessment.

WGNAS reassessed the effects of applying stricter criteria than $r^{2} \geq 0.2$ for inclusion of indicators in the FWI. As stricter criteria are used, the number of indicators included reduces rapidly. It was therefore concluded to keep the criterion of $r^{2} \geq 0.2$ in order to obtain a sufficient number of indicators to be able to use the FWI even in the event of one or more indicators being unavailable by the time the FWI is applied each year. The $r^{2}$ value of 0.2 corresponds to a value slightly lower than what is considered to be a "large" effect size ( $r=0.5, r^{2}=0.25$ ) by Cohen (1988). Although a criterion of $r^{2} \geq 0.2$ gives each indicator little predictive power alone (Prairie, 1996), the approach of using a suite of indicators is more similar to meta-analysis (Rosenthal, 1984) meaning that the
outcome of the FWI is not dependent on the result of one indicator in isolation, but rather on the combined performance of the indicator set.

The Faroes FWI was updated during the 2015 WGNAS meeting and an updated spreadsheet produced. Details are provided in Section 3.7 of the Working Group report (see above). In evaluating all the time-series, it was noted that the lower $12.5 \% \mathrm{CL}$, which is used to determine which indicator values are outside the $75 \% \mathrm{CI}$ on the lower side, was negative for some regression relationships for predicted PFA values in 2015 and 2016. Since this would invalidate the use of such indicators (they would not indicate that predicted PFA values are overestimates regardless of how small they are), an additional (fourth) criterion was established as a requirement for including timeseries in the FWI. This requires that the lower $12.5 \%$ confidence linit for an indicator time-series should be positive for any values of PFA included in the FWI.
The Working Group made further changes to the FWI in 2016 and 2018 (ICES, 2016, 2018). This followed the use of the FWI in January 2016 when the FWI signalled that the PFA of the Northern NEAC MSW stock complex was higher than forecast by the Working Group in 2015 and that a reassessment was necessary. However, in the catch advice provided in 2015 (ICES, 2015), it was the status of the Southern NEAC stock complexes which had resulted in a zero catch option for Faroes. As there was no indication from the FWI analysis that the forecast PFAs for these stocks had been underestimated, a change in the status of the Northern NEAC MSW stock complex alone would not have led to a change in the previous advice. To address this issue, an alternative FWI was developed, where only stock complexes that would be appropriate to changing the multi-year advice were included in the framework in the years between the provision of full catch advice. This revised FWI was adopted by NASCO in 2016 and again

As future catch advice could be determined by the status of stocks in any of the four stock complexes, it will be necessary to retain indicators for each of these in the FWI. Howeyer, in any year, it will only be necessary to apply the indicators from those stock complexes that could result in a change in the multi-year advice following a reassessment.


Figure 3.5.3.2.2. Output of the spreadsheet for the test of FWIs for NEAC for 2018 to 2022 based on the values of the indicators from 2018. The results of the NEAC FWI assessment in 2019 (based on indicator values for 2018) do not suggest that the PFA forecast for 2018 has been under-estimated. Therefore, the FWI Working Group concludes that no re-assessment of the existing management advice for the Faroes fishery is required from ICES in 2019..

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## Appendices to Stock Annex

Appendix 1 (a): Description of how catch and release mortality is incorporated in regional and national stock assessments


## Appendix 1 (b): Description of how unreported catch is incorporated in regional, national and international stock assessments



| Type of data | Collected under DCF | Available to WG | Reviewed and evaluated by WG | Used in current assessment models | Future plans |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| How to be filled | Yes/ | Yes/ | Yes/ | Yes/ | Keep as current DCF/ Impr | Free text |
|  | No/ | No/ | No/ | No/ | sampling intensity/ No need to |  |
|  | Partially | Partially | Partially | Partially used | , |  |
| Fleet capacity | No ** | No* | No | No | need to be collected | See 'Fishing gear and effort' |
| Fuel consumption | No ** | No * | No | No | need to be collected | Many salmon fisheries use unpowered vessels |
| Fishing gear and effort | Partially ** | Partially | Partially | Partially, but information requested by NASCO | Use for estimation of exploitation rates. <br> In mprove coverage and sampling intensity in DC-MAP | Data required for all relevant areas/fisheries |
| Landings | Partially ** | Yes | es |  | Improve coverage in DC-MAP | Data required on: catch in numbers and weights for recreational and commercial fisheries in rivers, estuaries and coastal waters. |
| Discards | No ** | $\mathrm{No}^{*}$ |  |  | No need to be collected | Not relevant to salmon except (historically) in Faroes fishery. NB: 'catch and release' fish are deliberately caught and so not classed as discards. |
| Recreational fisheries | Partially ** |  |  | Yes | Improve coverage in DC-MAP | Extent of DCF coverage unclear. <br> Complete catch data needed for all recreational fisheries (see 'Landings') |
| Catch \& Release | No ** | Partially | Partially | No - but data requested by NASCO | Include collection in DC-MAP | Data on numbers of fish caught and released required for all recreational fisheries |


| Type of data | Collected under DCF | Available to WG | Reviewed and evaluated by WG | Used in current assessment models | Future plans |
| :---: | :---: | :---: | :---: | :---: | :---: |
| CPUE dataseries | Partially ** | Partially | Partially | Partially | Improve sampling intensity DC-MAP |
| Age composition | Partially ** <br> Some ageing <br> based on fish <br> lengths or weights | Yes | Yes | Yes | Improve coverage and sampling intensity in |
| Wild/reared origin (scale reading) | No ** | Partially from other sources | Partially | Partially used information on farmed fish is requested by NASCO | Improve sampling intensity in DC-MAP |
| Length and weight-at-age | Partially ** | Partially | Yes | Yes - but some <br> ageing based on fish | prove sampling coverage in -MAP |
| Sex ratio | No ** | Yesfrom other sources | 11 y |  | Modify sampling intensity in DC-MAP |
| Maturity | Not known ** | No* |  |  | No need to be collected - all returning adults are mature |
| Fecundity | No ** | Yes | Partially |  | Include collection in DC-MAP |
| Data processing industry | No ** |  |  | No | No need to be collected |
| Juvenile surveys (Electrofishing) | Partially ** <br> but not requested <br> for Atlantic <br> salmon in DCF | Yes | Partially | Partially | Include collection in DC-MAP |



## Appendix 3: Input data for NEAC Pre Fishery Abundance analysis using Monte Carlo simulation

## Finland

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).

| $\begin{aligned} & \text { む̃ } \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 8422 | 8538 | 35.0 | 10.0 | 35.0 | 10.0 | 50.0 | 10.0 |  | 15.0 |
| 1972 | 32789 | 8950 | 35.0 | 10.0 | 35.0 | 10.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1973 | 15261 | 14402 | 35.0 | 10.0 | 35.0 | 10.0 |  | 10.0 | 55.0 | 15.0 |
| 1974 | 21057 | 24508 | 35.0 | 10.0 | 35.0 | 10.0 | 50. | 10.0 | 5.0 | 15.0 |
| 1975 | 25242 | 31347 | 35.0 | 10.0 | 35.0 | 10.0 | 50.0 | 0.0 | 55.0 | 15.0 |
| 1976 | 23000 | 24561 | 35.0 | 10.0 | 35.0 | 10.0 | 50.0 | 8 | 55.0 | 15.0 |
| 1977 | 12958 | 17035 | 35.0 | 10.0 | 35.0 | 10.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1978 | 12338 | 8670 | 35.0 | 10.0 | 35.0 | 10.0 | 50.0 | 10.0 | 55.0 | 15.0 |
| 1979 | 11071 | 7078 | 35.0 | 10.0 | 35.0 | 10.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1980 | 10097 | 7994 | 25.0 | 10.0 | 25.0 | 10.0 | 5 | 10.0 | 45.0 | 15.0 |
| 1981 | 9049 | 9476 | 25.0 | 10.0 |  | 10.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1982 | 5379 | 12628 | 25.0 | 10.0 | 25. | 10.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1983 | 13156 | 14013 | 25.0 | 10.0 | 25.0 | 10.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1984 | 14371 | 11718 | 25.0 | 10.0 |  | 10.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1985 | 19058 | 11299 | 25.0 | 10.0 | 25.0 | 10.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1986 | 15005 | 9320 | 25 | 10.0 | 25. | 10.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1987 | 18151 | 12208 | 25.0 | 10.0 | 25.4 | 10.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1988 | 10676 | 8631 | 25.0 | 10.0 | 5.0 | 10.0 | 50.0 | 10.0 | 45.0 | 15.0 |
| 1989 | 27956 | 1033 | 25.0 | 0 | 25.0 | 10.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1990 | 27955 | 1423 | 25 | 10.0 | 25.0 | 10.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1991 | 27513 | 15287 | 5.0 | 10.0 | 25.0 | 10.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1992 | 38843 | 1482 | 25.0 | 10.0 | 25.0 | 10.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1993 | 6195 | 15512 | 25.0 | 10.0 | 25.0 | 10.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1994 | 14555 | 4621 | 25.0 | 10.0 | 25.0 | 10.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 199 | 14525 | 9625 | 25.0 | 10.0 | 25.0 | 10.0 | 60.0 | 10.0 | 55.0 | 15.0 |
| 1996 |  | 8079 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 50.0 | 15.0 |
| 1997 | 8621 | 976 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 50.0 | 15.0 |
| 199 | 23 | 37 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 50.0 | 15.0 |
| 1999 | 37495 | 1071 | 25.0 | 10.0 | 25.0 | 10.0 | 60.0 | 10.0 | 50.0 | 10.0 |
|  | 40730 | 21088 | 25.0 | 10.0 | 25.0 | 10.0 | 60.0 | 10.0 | 50.0 | 10.0 |
| 2001 | 29501 | 28112 | 25.0 | 10.0 | 25.0 | 10.0 | 60.0 | 10.0 | 55.0 | 10.0 |
| 2002 | 16721 | 24642 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 003 | 16497 | 17751 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 004 | 7002 | 8062 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2005 | 15366 | 6685 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2006 | 26916 | 10533 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2007 | 7862 | 15269 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2008 | 8481 | 15355 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2009 | 15042 | 6587 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2010 | 12085 | 10590 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2011 | 13727 | 8152 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2012 | 23764 | 9851 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2013 | 13724 | 9494 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2014 | 19495 | 10302 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2015 | 12127 | 9905 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2016 | 9470 | 10584 | 25.0 | 10.0 | 25.0 | 10.0 | 55.0 | 10.0 | 55.0 | 10.0 |
| 2017 | 4676 | 6645 | 25.0 | 10.0 | 25.0 | 10.0 | 50.0 | 10.0 | 50.0 | 10.0 |
| 2018 | 11808 | 4078 | 25.0 | 10.0 | 25.0 | 19.0 | 50.0 | 10.0 | 50.0 | 10.0 |

## France

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Iceland (South and West)

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Iceland (North and East)

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


Ireland
Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).

| $\begin{gathered} \text { 厄્రే } \\ \hline \end{gathered}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1SW salmon in closed rivers |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 409965 | 46594 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |
| 1972 | 437089 | 49863 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1973 | 476131 | 54008 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 7.5 |  |  |  |  |  |  |  |
| 1974 | 542124 | 60976 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 |  | 12.5 |  |  |  |  |  |  |
| 1975 | 598524 | 68260 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 |  | 2.5 |  |  |  |  |  |  |
| 1976 | 407018 | 47358 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |
| 1977 | 351745 | 41256 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |
| 1978 | 307569 | 35708 | 37.5 | 7.5 | 37.5 | 7.5 |  | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |
| 1979 | 282700 | 32144 | 37.5 | 7.5 | 37.5 | 7.5 | 62. | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |
| 1980 | 215116 | 35447 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |
| 1981 | 137366 | 26101 | 37.5 | 7.5 | 37.5 |  |  |  | 47.5 | 12.5 |  |  |  |  |  |  |
| 1982 | 269847 | 11754 | 37.5 | 7.5 | 37.5 |  | 1.9 | 10.8 | 36.7 | 8.3 |  |  |  |  |  |  |
| 1983 | 437751 | 26479 | 37.5 | 7.5 | 37.5 |  | 66.1 | 9.9 | 40.1 | 7.5 |  |  |  |  |  |  |
| 1984 | 224872 | 20685 | 37.5 | 7.5 | 37.5 |  | 64.6 | 9.7 | 43.5 | 6.5 |  |  |  |  |  |  |
| 1985 | 430315 | 18830 | 37.5 | 7.5 | . | . 5 | 74.6 | 11.2 | 36.1 | 3.4 |  |  |  |  |  |  |
| 1986 | 443701 | 27111 | 37.5 | 7.5 | 37.5 |  | 68.7 | 10.3 | 46.0 | 9.0 |  |  |  |  |  |  |
| 1987 | 324709 | 26301 | 30.0 | 10.0 | 30.0 | 10.0 | 69.8 | 10.5 | 32.2 | 4.7 |  |  |  |  |  |  |
| 1988 | 391475 | 22067 | 30.0 | 10.0 |  |  | 62.0 | 9.3 | 37.4 | 5.6 |  |  |  |  |  |  |
| 1989 | 297797 | 25447 | 30.0 | 10.0 |  |  | 65.7 | 9.9 | 47.2 | 8.8 |  |  |  |  |  |  |
| 1990 | 172098 | 15549 | 30.0 | 10.0 | 30.0 | 0.0 | 60.7 | 9.1 | 59.9 | 6.1 |  |  |  |  |  |  |
| 1991 | 120408 | 10334 | 30.0 | 10.0 | 30.0 |  | 59.5 | 8.9 | 26.5 | 3.5 |  |  |  |  |  |  |
| $1992$ | 182255 | 15456 | 30.0 | 10. |  | 10.0 | 62.1 | 9.3 | 51.5 | 3.8 |  |  |  |  |  |  |
| 1993 | 150274 | 13156 | 25.0 |  | 25.0 | 10.0 | 58.6 | 8.8 | 42.0 | 18.0 |  |  |  |  |  |  |
| 1994 | 234126 | 20506 | 25.0 |  | 25.0 | 10.0 | 71.4 | 10.7 | 40.5 | 2.5 |  |  |  |  |  |  |
| 1995 | 232480 | 20454 | 25.0 | 10.0 | 25.0 | 10.0 | 63.5 | 9.5 | 41.8 | 1.2 |  |  |  |  |  |  |
| 1996 | 203920 | 18021 | 25.0 | 10.0 | 25.0 | 10.0 | 59.9 | 9.0 | 55.1 | $3.2$ |  |  |  |  |  |  |
| $1997$ | 170774 | 14724 | 25.0 | 10.0 | 15.0 | 5.0 | 50.1 | 7.5 | 30.8 | 12.2 |  |  |  |  |  |  |
| 1998 | 191868 | 17269 | 25.0 | 10.0 | 15.0 | 5.0 | 53.7 | 8.1 | 61.9 | 1.4 |  |  |  |  |  |  |
| 1999 | 158818 | 14801 | 25.0 | 10.0 | 15.0 | 5.0 | 47.8 | 7.2 | 34.1 | 18.1 |  |  |  |  |  |  |
| 2000 | 199827 | 16848 | 25.0 | 10.0 | 15.0 | 5.0 | 43.2 | 6.5 | 31.0 | 4.5 |  |  |  |  |  |  |
| $2001$ | 218715 | 18436 | 7.5 | 2.5 | 7.5 | 2.5 | 48.0 | 7.2 | 35.0 | 8.0 |  |  |  |  |  |  |
| 2002 | 198719 | 16702 | 7.5 | 2.5 | 7.5 | 2.5 | 49.9 | 7.5 | 27.5 | 7.5 |  |  |  |  |  |  |
| 2003 | 161270 | 13745 | 7.5 | 2.5 | 7.5 | 2.5 | 41.3 | 6.2 | 21.5 | 5.5 |  |  |  |  |  |  |


| $\begin{gathered} \text { だ } \\ \text { D } \end{gathered}$ |  |  |  |  |  |  |  |  |  |  | uoujes MSL Чフৃет łəu paлер |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 142251 | 12299 | 7.5 | 2.5 | 7.5 | 2.5 | 49.5 | 7.5 | 35.0 |  |  |  |  |  |  |  |  |  |
| 2005 | 127371 | 10716 | 7.5 | 2.5 | 7.5 | 2.5 | 44.5 | 6.5 | 23.5 | 3.5 |  |  |  |  |  |  |  |  |
| 2006 | 101938 | 9740 | 7.5 | 2.5 | 7.5 | 2.5 | 46.5 | 6.5 | 29.5 | 13.5 |  |  |  |  |  |  |  |  |
| 2007 | 17863 | 2867 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 8177 | 666 | 12137 | 988 | 0 | 0 | 24433 | 158 |
| 2008 | 31843 | 3935 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 8233 | 670 | 12071 | 1492 | 0 | 0 | 23259 | 213 |
| 2009 | 24268 | 4675 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 6248 | 509 | 9812 | 1610 | 0 | 0 | 30008 | 1873 |
| 2010 | 32981 | 4497 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 |  | 13093 | 1066 | 13325 | 1817 | 0 | 0 | 30605 | 616 |
| 2011 | 28105 | 4889 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 |  | 9.1 | 11071 | 902 | 11031 | 1657 | 0 | 0 | 28504 | 765 |
| 2012 | 29979 | 4197 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23. | 9.1 | 9542 | 777 | 10429 | 1463 | 0 | 0 | 24517 | 1213 |
| 2013 | 24029 | 4831 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 |  |  | 9.1 | 13378 | 747 | 8821 | 1861 | 0 | 0 | 23836 | 1250 |
| 2014 | 13787 | 4063 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 9173 | 397 | 5107 | 1430 | 0 | 0 | 20110 | 1210 |
| 2015 | 20835 | 4272 | 7.5 | 2.5 | 7.5 | 2.5 | 15.5 | 8.4 | 23.9 | 9.1 | 7396 | 295 | 7810 | 1573 | 0 | 0 | 25834 | 1134 |
| 2016 | 21619 | 3918 | 7.5 | 2.5 | 7.5 | 2.5 |  | 8.4 | 23.9 | 9.1 | 5755 | 1162 | 9413 | 1522 | 0 | 0 | 23953 | 1657 |
| 2017 | 23940 | 3782 | 7.5 | 2.5 | 7.5 | 2.5 | 15. | 8.4 | 23.9 | 9.1 | 5892 | 791 | 10977 | 1586 | 0 | 0 | 22590 | 1033 |
| 2018 | 18993 | 4254 | 7.5 | 2.5 | 7.5 | 2.5 |  |  | 23.9 | 9.1 | 5134 | 421 | 6027 | 1350 | 0 | 0 | 22427 | 1184 |

## Norway (South-east)

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Norway (South-west)

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Mid-Norway

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Norway North

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Russia (Archangelsk and Karelia)

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Russia (Kola Peninsula: Barents Sea Basin)

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Russia (Kola Peninsula: White Sea Basin)

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Russia (Pechora River)

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Sweden

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## UK (England and Wales)

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


| $\begin{array}{r} \text { む̈ } \\ \hline \end{array}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 19914 | 0.78 |  | 4894 | 2197 | 0.78 | 17.7 | 4.4 | 17.7 |  | 17.9 |  | 8.4 | 4.2 | 17.1 |  | 0.75 | 0.50 |
| 2008 | 19036 | 0.76 |  | 3649 | 2592 | 0.76 | 17.8 | 4.4 | 17.8 |  | 17. |  | 8.2 | 4.1 | 17.1 |  | 0.75 | 0.50 |
| 2009 | 13910 | 0.72 |  | 2590 | 2805 | 0.72 | 11.4 | 2.9 | 11.4 |  |  | 8.7 | 8.2 | 4.1 | 7.4 |  | 0.75 | 0.50 |
| 2010 | 32695 | 0.78 |  | 12214 | 7768 | 0.78 | 10.8 | 2.7 | 10.8 | 2.7 | 17.5 |  | 8.0 | 4.0 | 7.4 |  | 0.75 | 0.50 |
| 2011 | 34575 | 0.57 |  | 14915 | 9233 | 0.57 | 10.1 | 2.5 | 10.1 |  | 20.8 | 10.0 | 10.2 | 5.1 | 7.4 |  | 0.64 | 0.37 |
| 2012 | 14926 | 0.50 |  | 3571 | 3705 | 0.50 | 11.2 | 2.8 |  | 2.8 | 6.8 | 8.4 | 8.0 | 4.0 | 7.4 |  | 0.62 | 0.37 |
| 2013 | 22608 | 0.58 |  | 7964 | 8679 | 0.58 | 9.5 | 2.4 |  | 2.4 |  | 8.7 | 8.5 | 4.3 | 7.4 |  | 0.63 | 0.37 |
| 2014 | 14219 | 0.54 |  | 6974 | 3826 | 0.54 | 9.3 | 2.3 |  | 2.3 | 15.8 | 7.9 | 8.0 | 4.0 | 7.4 |  | 0.64 | 0.37 |
| 2015 | 19262 | 0.47 |  | 9233 | 6657 | 0.47 | 12.9 | 3.2 |  | 3.2 | 15.2 | 7.6 | 7.7 | 3.9 | 7.4 |  | 0.64 | 0.37 |
| 2016 | 22494 | 0.42 |  | 10811 | 7956 | 0.42 | 12.0 | 3.0 |  | 3.0 | 14.8 | 7.4 | 7.5 | 3.7 | 7.4 |  | 0.63 | 0.37 |
| 2017 | 12195 | 0.40 |  | 5095 | 3975 | 0.40 | 13.8 |  | 13 |  | 11.7 | 5.8 | 5.8 | 2.9 | 7.4 |  | 0.63 | 0.38 |
| 2018 | 11640 | 0.45 |  | 4059 | 5839 | 0.45 | 12.4 |  |  | 3.1 | 9.4 | 4.7 | 4.9 | 2.4 | 7.4 |  | 0.63 | 0.37 |

## UK (Northern Ireland)-Foyle Fisheries Area

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).

| $\begin{gathered} \text { む̈ } \\ \hline \end{gathered}$ |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 78037 | 5874 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 800 | 5.0 |
| 1972 | 64663 | 4867 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 0 |
| 1973 | 57469 | 4326 |  |  | 21.5 | 11.5 | 21.5 | 11 | 80.0 | 5.0 |
| 1974 | 72587 | 5464 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 |
| 1975 | 51061 | 3843 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 0 |
| 1976 | 36206 | 2725 |  |  | 21.5 | 11.5 | 21.5 | 11.5 |  | 5.0 |
| 1977 | 36510 | 2748 |  |  | 21.5 | 11.5 | 1.5 | 1. | 80. | 5.0 |
| 1978 | 44557 | 3354 |  |  | 21.5 | 11.5 | 21. | 1.5 | 80.0 | 5.0 |
| 1979 | 34413 | 2590 |  |  | 21.5 | 11.5 | 21 | 11.5 | 80.0 | 5.0 |
| 1980 | 45777 | 3446 |  |  | 21.5 | 11. | 21.5 | 11.5 | 80.0 | 5.0 |
| 1981 | 32346 | 2435 |  |  | 21.5 | 11.5 | 21.5 |  | 80.0 | 5.0 |
| 1982 | 55946 | 4211 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 |
| 1983 | 77424 | 5828 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 |
| 1984 | 27465 | 2067 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 |
| 1985 | 37685 | 2836 |  |  | 21.5 |  |  | 11.5 | 80.0 | 5.0 |
| 1986 | 43109 | 3245 |  |  |  | 11.5 | 21.5 | 11.5 | 80.0 | 5.0 |
| 1987 | 17189 | 1294 |  |  | 21.5 | 11 | 21.5 | 11.5 | 69.0 | 7.0 |
| 1988 | 43974 | 3310 |  |  |  | 11.5 | 21.5 | 11.5 | 64.5 | 6.5 |
| 1989 | 60288 | 4538 |  |  |  | 13.5 | 23.5 | 13.5 | 89.0 | 9.0 |
| 1990 | 39875 | 3001 |  |  |  | 3.5 | 13.5 | 3.5 | 62.0 | 6.0 |
| 1991 | 21709 | 1634 |  |  | 13.5 | 3.5 | 13.5 | 3.5 | 64.5 | 6.5 |
| 1992 | 39299 | 2958 |  |  | 16.5 | 6.5 | 16.5 | 6.5 | 56.0 | 6.0 |
| 1993 | 35366 | 2662 |  |  | 3.5 | 3.5 | 13.5 | 3.5 | 41.0 | 4.0 |
| 1994 | 36144 | 2720 |  |  | 19.0 | 9.0 | 19.0 | 9.0 | 70.0 | 7.0 |
| 1995 | 33398 | 514 |  |  | 13.5 | 3.5 | 13.5 | 3.5 | 67.0 | 7.0 |
| 1996 | 28406 |  |  |  | 15.0 | 5.0 | 15.0 | 5.0 | 57.0 | 10.0 |
| 1997 | 40886 | 308 |  |  | 10.0 | 5.0 | 10.0 | 5.0 | 60.0 | 10.0 |
| 1998 | 7154 | 2797 |  |  | 10.0 | 5.0 | 10.0 | 5.0 | 25.0 | 5.0 |
| 1999 | 1660 | 630 |  |  | 10.0 | 5.0 | 10.0 | 5.0 | 63.0 | 5.0 |
| 2000 | 30385 | 287 |  |  | 10.0 | 5.0 | 10.0 | 5.0 | 58.0 | 5.0 |
| 2001 |  | 1608 |  |  | 5.0 | 5.0 | 5.0 | 5.0 | 50.0 | 5.0 |
| 2002 | 37914 |  | 9163 | 690 | 2.5 | 2.5 | 2.5 | 2.5 | 15.0 | 3.0 |
| 2003 | 30 | 291 | 4576 | 344 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 |
| 2004 | 20730 | 1560 | 4570 | 344 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 |
|  | 23746 | 1787 | 7079 | 533 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 |
| 2006 | 11324 | 852 | 4886 | 368 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 |
| 2007 | 5050 | 322 | 9530 | 608 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 |
| 2008 | 3880 | 292 | 4755 | 304 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 |
| 2009 | 1743 | 194 | 3640 | 405 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 |
| 2010 | 0 | 0 | 3488 | 388 | 0.5 | 0.5 | 0.5 | 0.5 | 15.0 | 3.0 |
| 2011 | 0 | 0 | 2276 | 759 | 1.0 | 1.0 | 1.0 | 1.0 | 15.0 | 5.0 |
| 2012 | 0 | 0 | 4781 | 1594 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 7.5 |
| 2013 | 0 | 0 | 5030 | 498 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 7.5 |
| 2014 | 0 | 0 | 2029 | 225 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 7.5 |
| 2015 | 0 | 0 | 1998 | 250 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 7.5 |
| 2016 | 0 | 0 | 3192 | 355 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 7.5 |
| 2017 | 0 | 0 | 3511 | 347 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 7.5 |
| 2018 | 0 | 0 | 3008 | 298 | 1.0 | 1.0 | 1.0 | 1.0 | 10.0 | 7.5 |

## UK (Northern Ireland)-DAERA area

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).

| $\begin{aligned} & \text { む̈ } \\ & \text { た } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  | Estimated exploitation rate (\%) of MSW salmon |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 35506 | 2673 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |  | 0 | 0 | 0 |  |  |  |  |
| 1972 | 34550 | 2601 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |  | 0 | 0 | 0 |  |  |  |  |
| 1973 | 29229 | 2200 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 |  | 50 |  | 0 | 0 | 0 | 0 |  |  |  |  |
| 1974 | 22307 | 1679 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 |  | 50 |  | 0 | 0 | 0 | 0 |  |  |  |  |
| 1975 | 26701 | 2010 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 |  | 0 | 0 | 0 | 0 |  |  |  |  |
| 1976 | 17886 | 1346 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 |  |  | 0 | 0 | 0 | 0 |  |  |  |  |
| 1977 | 16778 | 1263 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 |  | 5 | 0 | 0 | 0 | 0 |  |  |  |  |
| 1978 | 24857 | 1871 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 | 0 | 0 | 0 | 0 |  |  |  |  |
| 1979 | 14323 | 1078 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 | 0 | 0 | 0 | 0 |  |  |  |  |
| 1980 | 15967 | 1202 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 | 0 | 0 | 0 | 0 |  |  |  |  |
| 1981 | 15994 | 1204 |  |  | 21.5 | 11.5 | 21.5 | , | 80 |  | 50 | 5 | 0 | 0 | 0 | 0 |  |  |  |  |
| 1982 | 14068 | 1059 |  |  | 21.5 | 11.5 | 21.5 | 11.5 |  |  | 50 | 5 | 0 | 0 | 0 | 0 |  |  |  |  |
| 1983 | 20845 | 1569 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 |  | 50 | 5 | 0 | 0 | 0 | 0 |  |  |  |  |
| 1984 | 11109 | 836 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 | 0 | 0 | 0 | 0 |  |  |  |  |
| 1985 | 12369 | 931 |  |  | 21.5 | 11.5 | 21.5 |  | 80 | 5 | 50 | 5 | 0 | 0 | 0 | 0 |  |  |  |  |
| 1986 | 13160 | 991 |  |  | 21.5 | 11.5 | 21.5 |  | 80 | 5 | 50 | 5 | 0 | 0 | 0 | 0 |  |  |  |  |
| 1987 | 9240 | 695 |  |  | 21.5 |  | 21.5 |  |  |  |  |  | 2530 | 83 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 1988 | 14320 | 1078 |  |  | 21.5 |  | . | 1.5 |  |  |  |  | 2832 | 96 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 1989 | 15081 | 1135 |  |  | 23.5 | 13 | 23.5 |  |  |  |  |  | 1029 | 82 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 1990 | 9499 | 715 |  |  | 13.5 |  | 13.5 | 3.5 |  |  |  |  | 1850 | 87 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 1991 | 6987 | 526 |  |  | 13.5 |  |  | 3.5 |  |  |  |  | 2341 | 87 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 1992 | 9346 | 703 |  |  | 16.5 | 6.5 | 16.5 | 6.5 |  |  |  |  | 2546 | 84 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 1993 | 7906 | 595 |  |  | 13.5 |  | -13.5 | 3.5 |  |  |  |  | 3235 | 93 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 1994 | 11206 | 843 |  |  |  | 9.0 | 19.0 | 9.0 |  |  |  |  | 2010 | 88 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 1995 | 11637 | 876 |  |  |  | 3.5 | 13.5 | 3.5 |  |  |  |  | 1521 | 92 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 1996 | 10383 | 781 |  |  |  | 5.0 | 15.0 | 5.0 |  |  |  |  | 1097 | 87 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 1997 | 10479 | 789 |  |  | 10.0 | 5.0 | 10.0 | 5.0 |  |  |  |  | 1677 | 85 | 6541 | 85 | 67.0 | 5.0 | 61.0 | 5.0 |


| $\begin{aligned} & \text { テ̈ } \\ & \text { ジ } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 9375 | 706 |  |  | 10.0 | 5.0 | 10.0 | 5.0 |  |  |  |  | 2995 | 95 |  | 95 | 67.0 | 5.0 | 61.0 | 5.0 |
| 1999 | 9011 | 678 |  |  | 10.0 | 5.0 | 10.0 | 5.0 |  |  |  |  | 977 | 90 | 3599 | 90 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2000 | 10598 | 798 |  |  | 10.0 | 5.0 | 10.0 | 5.0 |  |  |  |  |  |  | 5979 | 91 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2001 | 8104 | 610 |  |  | 5.0 | 5.0 | 5.0 | 5.0 |  |  |  |  |  |  | 5771 | 97 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2002 | 3315 | 249 | 2218 | 167 | 2.5 | 2.5 | 2.5 | 2.5 |  |  |  |  |  |  | 5037 | 95 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2003 | 2236 | 168 | 1884 | 141 | 2.5 | 2.5 | 2.5 | 2.5 |  |  |  |  |  | 96 | 4147 | 96 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2004 | 2411 | 181 | 3053 | 230 | 0.5 | 0.5 | 0.5 | 0.5 |  |  |  |  | 878 | 92 | 9050 | 92 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2005 | 3012 | 227 | 1791 | 135 | 0.5 | 0.5 | 0.5 | 0.5 |  |  |  |  | 1151 | 91 | 6609 | 91 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2006 | 2288 | 172 | 1289 | 97 | 0.5 | 0.5 | 0.5 | 0.5 |  |  |  |  | 1074 | 87 | 7410 | 87 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2007 | 2533 | 162 | 2427 | 155 | 0.5 | 0.5 | 0.5 | 0.5 |  |  |  |  | 2584 | 94 | 7008 | 94 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2008 | 1825 | 116 | 2444 | 156 | 0.5 | 0.5 | 0.5 | 0.5 |  |  |  |  | 1712 | 90 | 0 | 0 | 10.0 | 2.0 | 9.5 | 2.0 |
| 2009 | 1383 | 154 | 1457 | 162 | 0.5 | 0.5 | 0.5 | 0.5 |  |  |  |  | 726 | 83 | 3838 | 83 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2010 | 1723 | 191 | 1327 | 147 | 0.5 | 0.5 | 0.5 | 0.5 |  |  |  |  | 1045 | 78 | 6426 | 70 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2011 | 857 | 285 | 1132 | 378 | 1.0 | 1.0 | 1.0 | 1.0 |  |  |  |  | 649 | 73 | 6130 | 76 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2012 | 15 | 5 | 263 | 87 | 1.0 | 1.0 | 1.0 | 1.0 |  |  |  |  | 926 | 74 | 5319 | 71 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2013 | 9 | 1 | 46 | 5 | 1.0 | 1.0 | 1.0 | 1.0 |  |  |  |  | 1644 | 92 | 5866 | 91 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2014 | 0 | 0 | 143 | 40 | 2.5 | 2.5 | 2.5 |  |  |  |  |  | 963 | 76 | 4335 | 91 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2015 | 0 | 0 | 20 | 6 | 2.5 | 2.5 | 2.5 | 2.5 |  |  |  |  | 1005 | 83 | 6235 | 86 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2016 | 0 | 0 | 112 | 12 | 2.5 | 2.5 |  | 2.5 |  |  |  |  | 2166 | 85 | 15936 | 86 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2017 | 0 | 0 | 82 | 15 | 2.5 | 2.5 |  | 2.5 |  |  |  |  | 912 | 88 | 8433 | 81 | 67.0 | 5.0 | 61.0 | 5.0 |
| 2018 | 0 | 0 | 23 | 102 | 2.5 | 2.5 |  |  |  |  |  |  | 728 | 81 | 8853 | 83 | 67.0 | 5.0 | 61.0 | 5.0 |

## UK（Scotland）－East

Annual input data for NEAC PFA run－reconstruction \＆NCL models（uncertainty values define uniform distribution around estimates used in Monte Carlo simulation）．

| غ゙ ジ |  |  | $\begin{aligned} & \text { Declared net catch 1SW } \\ & \text { salmon } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 4581 | 0 | 212292 | 0.012931 | 0.001781 | 0.10 | 0.05 | 34192 | 0 | 101338 |  | 0.027497 | 0.10 | 0.05 | 4708 | 7115 |
| 1972 | 4672 | 0 | 215434 | 0.011596 | 0.001765 | 0.10 | 0.05 | 45211 |  | 138664 | 0.149076 | 0.026541 | 0.10 | 0.05 | 4803 | 7270 |
| 1973 | 5277 | 0 | 254496 | 0.010981 | 0.001664 | 0.10 | 0.05 | 49569 |  | 155257 | 0.161510 | 0.026943 | 0.10 | 0.05 | 4858 | 7391 |
| 1974 | 5971 | 0 | 239453 | 0.013363 | 0.001993 | 0.10 | 0.05 | 41764 |  | 117195 | 0.188422 | 0.030035 | 0.10 | 0.05 | 4860 | 7460 |
| 1975 | 4718 | 0 | 177222 | 0.012455 | 0.001999 | 0.10 | 0.05 | 54153 | 0 | 126675 | 0.175011 | 0.033018 | 0.10 | 0.05 | 4808 | 7473 |
| 1976 | 5287 | 0 | 144782 | 0.017891 | 0.002814 | 0.10 | 0.05 | 3377 | 0 | 58409 | 0.193866 | 0.040357 | 0.10 | 0.05 | 4709 | 7435 |
| 1977 | 6648 | 0 | 147658 | 0.018602 | 0.003250 | 0.10 | 0.05 | 4941 |  | 69226 | 0.209021 | 0.045315 | 0.10 | 0.05 | 4580 | 7362 |
| 1978 | 7913 | 0 | 150946 | 0.021738 | 0.003827 | 0.10 | 0.05 | 59080 |  | 80683 | 0.173083 | 0.039986 | 0.10 | 0.05 | 4439 | 7275 |
| 1979 | 10760 | 0 | 150036 | 0.027987 | 0.004994 | 0.10 | 0.05 | 58124 | 0 | 58435 | 0.211189 | 0.049860 | 0.10 | 0.05 | 4308 | 7198 |
| 1980 | 7336 | 0 | 94329 | 0.030350 | 0.005359 | 0.10 | 0.05 | 52184 | 0 | 103462 | 0.148937 | 0.032306 | 0.10 | 0.05 | 4203 | 7151 |
| 1981 | 8409 | 0 | 121281 | 0.026463 | 0.004721 | 0.10 | 0.05 | 42896 | 0 | 113787 | 0.150808 | 0.027926 | 0.10 | 0.05 | 4133 | 7143 |
| 1982 | 12417 | 0 | 162957 | 0.034158 | 0.005410 | 0.10 | 0.0 | 40398 | 0 | 72800 | 0.200489 | 0.038588 | 0.10 | 0.05 | 4105 | 7180 |
| 1983 | 9670 | 0 | 161173 | 0.024895 | 0.004187 | 0.10 | 0.05 | 4367 | 0 | 82433 | 0.204862 | 0.038318 | 0.10 | 0.05 | 4116 | 7255 |
| 1984 | 10557 | 0 | 165118 | 0.028930 | 0.004666 | 0.10 | 0.05 |  | 0 | 54508 | 0.195977 | 0.042160 | 0.10 | 0.05 | 4159 | 7354 |
| 1985 | 12427 | 0 | 120744 | 0.035858 | 0.006484 | 0.10 | 0.05 | 47258 | 0 | 47811 | 0.246188 | 0.054829 | 0.10 | 0.05 | 4223 | 7458 |
| 1986 | 11519 | 0 | 168773 | 0.027839 | 0.004834 |  | 0.05 | 48519 | 0 | 80135 | 0.203929 | 0.041271 | 0.10 | 0.05 | 4294 | 7546 |
| 1987 | 13710 | 0 | 125549 | 0.037471 | 0.006918 |  | 0.05 | 44326 | 0 | 44205 | 0.249639 | 0.056261 | 0.10 | 0.05 | 4357 | 7599 |
| 1988 | 19262 | 0 | 99358 | 0.046772 | 0.009537 | 0.10 | 0.05 | 53778 | 0 | 37389 | 0.312434 | 0.072704 | 0.10 | 0.05 | 4399 | 7608 |
| 1989 | 21251 | 0 | 121812 | 0.043918 | 0.009216 | 0.10 | 0.05 | 46689 | 0 | 38710 | 0.270677 | 0.062862 | 0.10 | 0.05 | 4413 | 7572 |
| 1990 | 13946 | 0 | 49406 | 0.045438 | 0.010452 | 0.10 | 0.05 | 42634 | 0 | 31340 | 0.242686 | 0.059998 | 0.10 | 0.05 | 4396 | 7500 |
| 1991 | 12544 | 0 | 41317 | 0.047342 | 0.010905 | 0.10 | 0.05 | 37497 | 0 | 16196 | 0.254299 | 0.066711 | 0.10 | 0.05 | 4353 | 7408 |
| 1992 | 21544 | 0 | 58362 | 0.05955 | 0.013 |  | 0.05 | 45154 | 0 | 22825 | 0.321138 | 0.077875 | 0.10 | 0.05 | 4291 | 7314 |
| 1993 | 22888 | 0 | 50525 | 059052 | 0.014162 | 10 | 0.05 | 43860 | 0 | 16638 | 0.304362 | 0.077632 | 0.10 | 0.05 | 4222 | 7234 |
| 1994 | 19418 | 1295 | 61011 | 0.051883 | 0.011796 | 0.10 | 0.05 | 45550 | 4634 | 27208 | 0.284761 | 0.070242 | 0.10 | 0.05 | 4155 | 7179 |
| 1995 | 18650 | 2217 | 54370 | 0.052460 | 0.011841 | 0.10 | 0.05 | 45935 | 8267 | 23120 | 0.262846 | 0.068156 | 0.10 | 0.05 | 4096 | 7151 |
| 1996 | 16869 | 1716 | 39758 | 0.056972 | 0.013940 | 0.10 | 0.05 | 34573 | 7402 | 15792 | 0.238648 | 0.064811 | 0.10 | 0.05 | 4047 | 7145 |
| 1997 | 14445 | 2228 | 23003 | 0.061842 | 0.015143 | 0.10 | 0.05 | 28128 | 7400 | 6722 | 0.278229 | 0.075232 | 0.10 | 0.05 | 4009 | 7150 |
| 1998 | 22797 | 4337 | 22155 | 0.088869 | 0.021884 | 0.10 | 0.05 | 27439 | 7721 | 4792 | 0.339571 | 0.092307 | 0.10 | 0.05 | 3978 | 7151 |
| 1999 | 10113 | 3020 | 10794 | 0.066303 | 0.016279 | 0.10 | 0.05 | 22140 | 10185 | 4871 | 0.273680 | 0.075183 | 0.10 | 0.05 | 3949 | 7132 |


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| 2000 | 14143 | 5967 | 22728 | 0.066682 | 0.016645 | 0.10 | 0.05 | 22630 | 12306 | 8650 | . 290559 |  | 0.10 | 0.05 | 3916 | 7082 |
| 2001 | 14900 | 7235 | 21746 | 0.068658 | 0.017196 | 0.10 | 0.05 | 22571 | 16689 | 7899 | 0.240691 |  | 0.10 | 0.05 | 3877 | 6999 |
| 2002 | 11315 | 6520 | 15301 | 0.077311 | 0.018888 | 0.10 | 0.05 | 16141 | 13830 | 55 | 0.266058 | 0.072265 | 0.10 | 0.05 | 3829 | 6886 |
| 2003 | 6823 | 7651 | 19048 | 0.056321 | 0.014029 | 0.10 | 0.05 | 12827 | 18255 | 11443 | 220976 | 0.059187 | 0.10 | 0.05 | 3776 | 6758 |
| 2004 | 14543 | 12722 | 17124 | 0.079619 | 0.019138 | 0.10 | 0.05 | 23284 | 27819 | 7489 | 267545 | 0.073272 | 0.10 | 0.05 | 3719 | 6630 |
| 2005 | 13437 | 12633 | 18160 | 0.079096 | 0.019785 | 0.10 | 0.05 | 17424 | 26039 | 6252 | . 238349 | 0.064738 | 0.10 | 0.05 | 3665 | 6521 |
| 2006 | 15649 | 16344 | 15090 | 0.102853 | 0.025555 | 0.10 | 0.05 | 16298 | 25 | 6656 | 0.1 | 0.051475 | 0.10 | 0.05 | 3618 | 6445 |
| 2007 | 13699 | 18807 | 12316 | 0.107991 | 0.026838 | 0.10 | 0.05 | 14907 | 269 | 4537 | 0.233134 | 0.063741 | 0.10 | 0.05 | 3580 | 6408 |
| 2008 | 10050 | 14458 | 8536 | 0.098004 | 0.024328 | 0.10 | 0.05 | 15557 | 312 | 5200 | 0.194359 | 0.053841 | 0.10 | 0.05 | 3554 | 6412 |
| 2009 | 8302 | 13334 | 6561 | 0.111169 | 0.027118 | 0.10 | 0.05 | 10641 | 287 | 4401 | 0.197725 | 0.052949 | 0.10 | 0.05 | 3537 | 6448 |
| 2010 | 12971 | 28380 | 15250 | 0.117455 | 0.028672 | 0.10 | 0.05 | 134 | 41199 |  | 0.210913 | 0.055822 | 0.10 | 0.05 | 3526 | 6502 |
| 2011 | 6260 | 11877 | 6281 | 0.099392 | 0.024222 | 0.10 | 0.05 | 123 | 44169 | 12008 | 0.173659 | 0.045921 | 0.10 | 0.05 | 3515 | 6554 |
| 2012 | 7347 | 16755 | 8785 | 0.103236 | 0.026107 | 0.10 | 0.05 | 1012 |  | 6100 | 0.184276 | 0.050846 | 0.10 | 0.05 | 3499 | 6583 |
| 2013 | 4274 | 12858 | 14126 | 0.087855 | 0.021572 | 0.10 | 0.05 | 6307 | 519 | 8594 | 0.169849 | 0.046160 | 0.10 | 0.05 | 3473 | 6572 |
| 2014 | 2517 | 8077 | 8407 | 0.099462 | 0.022969 | 0.10 | 0.05 | 3810 | 24008 | 7988 | 0.171343 | 0.044771 | 0.10 | 0.05 | 3435 | 6511 |
| 2015 | 3858 | 14642 | 8365 | 0.103699 | 0.025830 | 0.10 | 0.05 | 3589 | 26394 | 4203 | 0.154547 | 0.042580 | 0.10 | 0.05 | 3383 | 6403 |
| 2016 | 2600 | 12844 | 1441 | 0.092205 | 0.024621 | 0.10 |  | 2649 | 29790 | 1370 | 0.150420 | 0.043339 | 0.10 | 0.05 | 3322 | 6255 |
| 2017 | 2009 | 11143 | 908 | 0.093215 | 0.024292 | 0.10 |  | 2503 | 27259 | 1210 | 0.151827 | 0.043249 | 0.10 | 0.05 | 3254 | 6086 |
| 2018 | 1308 | 10429 | 2316 | 0.091281 | 0.024591 | 0.10 | 0.05 | 933 | 19438 | 1387 | 0.196434 | 0.055961 | 0.10 | 0.05 | 3184 | 5909 |

## UK (Scotland)-West

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

| 艺 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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| 1971 | 3497 | 0 | 41790 | 0.0338000 | 0.0059881 | 0.10 | 0.05 | 7255 | 0 | 18816 | 0.224202155 | 0.032722657 | 0.10 | 0.05 | 4708 | 7115 |
| 1972 | 2078 | 0 | 29280 | 0.0240317 | 0.0046632 | 0.10 | 0.05 | 7684 |  | 26464 | 0.185377258 | 0.024767867 | 0.10 | 0.05 | 4803 | 7270 |
| 1973 | 2495 | 0 | 30822 | 0.0238651 | 0.0049064 | 0.10 | 0.05 | 8965 |  | 24129 | 0.246207012 | 0.032074043 | 0.10 | 0.05 | 4858 | 7391 |
| 1974 | 3605 | 0 | 40387 | 0.0327150 | 0.0060787 | 0.10 | 0.05 | 8551 |  |  | 0.275836057 | 0.03486275 | 0.10 | 0.05 | 4860 | 7460 |
| 1975 | 2510 | 0 | 37707 | 0.0300866 | 0.0049060 | 0.10 | 0.05 |  |  | 20198 | 0.222944124 | 0.02882621 | 0.10 | 0.05 | 4808 | 7473 |
| 1976 | 2518 | 0 | 35736 | 0.0284048 | 0.0048811 | 0.10 | 0.05 |  |  | 15243 | 0.298263963 | 0.038435151 | 0.10 | 0.05 | 4709 | 7435 |
| 1977 | 2130 | 0 | 37609 | 0.0254128 | 0.0042048 | 0.10 | 0.05 |  |  | 13929 | 0.2586409 | 0.038597402 | 0.10 | 0.05 | 4580 | 7362 |
| 1978 | 3357 | 0 | 41985 | 0.0340762 | 0.0059303 | 0.10 | 0.05 |  | 0 | 16078 | 0.245306323 | 0.036359191 | 0.10 | 0.05 | 4439 | 7275 |
| 1979 | 4484 | 0 | 21800 | 0.0546500 | 0.0111746 | 0.10 | 0.05 | 7 | 0 | 8126 | 0.326599675 | 0.065220326 | 0.10 | 0.05 | 4308 | 7198 |
| 1980 | 3831 | 0 | 15823 | 0.0621272 | 0.0129164 | 0.10 | 0.05 | 7153 | 0 | 9729 | 0.233694125 | 0.046962651 | 0.10 | 0.05 | 4203 | 7151 |
| 1981 | 3863 | 0 | 16779 | 0.0536181 | 0.0115093 | 0.10 | . 05 | 8093 | 0 | 9803 | 0.22819773 | 0.049662238 | 0.10 | 0.05 | 4133 | 7143 |
| 1982 | 4422 | 0 | 27930 | 0.0506356 | 0.0097446 |  |  | 7517 | 0 | 7488 | 0.367043856 | 0.070526474 | 0.10 | 0.05 | 4105 | 7180 |
| 1983 | 4439 | 0 | 33967 | 0.0470398 | 0.0083839 | 10 | 0.05 | 8290 | 0 | 11488 | 0.370706596 | 0.056353803 | 0.10 | 0.05 | 4116 | 7255 |
| 1984 | 4968 | 0 | 32294 | 0.0473719 | 0.0100321 |  | 0.05 | 6769 | 0 | 9552 | 0.260598151 | 0.052638837 | 0.10 | 0.05 | 4159 | 7354 |
| 1985 | 5222 | 0 | 19374 | 0.0603708 | 0.0124453 |  | 0.05 | 11183 | 0 | 8313 | 0.480542092 | 0.09250515 | 0.10 | 0.05 | 4223 | 7458 |
| 1986 | 4200 | 0 | 18283 | 0.0440102 | 0.0100128 |  | 0.05 | 10723 | 0 | 8784 | 0.371853866 | 0.077795029 | 0.10 | 0.05 | 4294 | 7546 |
| 1987 | 4350 | 0 | 21038 | 0.0528866 | 0.0109220 |  | 0.05 | 8740 | 0 | 6706 | 0.513360196 | 0.094556632 | 0.10 | 0.05 | 4357 | 7599 |
| 1988 | 8547 | 0 | 21765 | 0.0690962 | 0151488 | 0.10 | 0.05 | 14901 | 0 | 6071 | 0.815724948 | 0.162020293 | 0.10 | 0.05 | 4399 | 7608 |
| 1989 | 8418 | 0 | 23370 | 0.0660006 | 8693 | 0.10 | 0.05 | 11649 | 0 | 6825 | 0.608859332 | 0.11965596 | 0.10 | 0.05 | 4413 | 7572 |
| 1990 | 5420 | 0 | 12309 | 0.0663459 | 0.015 |  | 0.05 | 9646 | 0 | 4296 | 0.609608749 | 0.129125105 | 0.10 | 0.05 | 4396 | 7500 |
| 1991 | 4770 | 0 | 14957 | 0.06377 | 13 | . 10 | 0.05 | 7639 | 0 | 3861 | 0.534661305 | 0.112494616 | 0.10 | 0.05 | 4353 | 7408 |
| 1992 | 6327 | 0 | 15443 | 0.0705496 | 015636 | 0.10 | 0.05 | 9872 | 0 | 4990 | 0.660809977 | 0.127470994 | 0.10 | 0.05 | 4291 | 7314 |
| 1993 | 5288 | 0 | 15816 | . 0639055 | 0.0143542 | 0.10 | 0.05 | 7441 | 0 | 3787 | 0.560810391 | 0.118734251 | 0.10 | 0.05 | 4222 | 7234 |
| 1994 | 4309 | 238 | 13925 | 66246 | 0.0129323 | 0.10 | 0.05 | 7713 | 428 | 4591 | 0.574756998 | 0.113552106 | 0.10 | 0.05 | 4155 | 7179 |
| 1995 | 4203 | 1086 | 12581 | 0.0776 | 0.0166713 | 0.10 | 0.05 | 5305 | 581 | 3828 | 0.500391392 | 0.096071777 | 0.10 | 0.05 | 4096 | 7151 |
| 1996 | 2909 | 566 | 6628 | 0.06772 | 0.0158645 | 0.10 | 0.05 | 4905 | 729 | 2558 | 0.611117207 | 0.129029339 | 0.10 | 0.05 | 4047 | 7145 |
| 1997 | 3319 | 581 | 5740 | 0.0798977 | 0.0188477 | 0.10 | 0.05 | 3907 | 735 | 1597 | 0.576929699 | 0.133031047 | 0.10 | 0.05 | 4009 | 7150 |
| 1998 | 4525 | 596 | 3844 | 0.0940776 | 0.0234844 | 0.10 | 0.05 | 5202 | 810 | 948 | 0.715678755 | 0.179929871 | 0.10 | 0.05 | 3978 | 7151 |


| $\begin{aligned} & \text { च゙ } \\ & \underset{\sim}{2} \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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| 1999 | 2556 | 834 | 1591 | 0.0951097 | 0.0240046 | 0.10 | 0.05 | 2881 | 810 |  | 0.55578466 | 39845771 | 0.10 | 0.05 | 3949 | 7132 |
| 2000 | 3590 | 1409 | 3384 | 0.0918593 | 0.0218781 | 0.10 | 0.05 | 4397 | 1390 |  | 2487837 | 0.198596617 | 0.10 | 0.05 | 3916 | 7082 |
| 2001 | 3673 | 2151 | 1930 | 0.0984072 | 0.0250378 | 0.10 | 0.05 | 3492 | 1649 | 69 | 41751016 | 0.1629531 | 0.10 | 0.05 | 3877 | 6999 |
| 2002 | 2747 | 1728 | 1944 | 0.0865669 | 0.0215752 | 0.10 | 0.05 | 3732 | 1980 | 816 | 838 | 0.176044603 | 0.10 | 0.05 | 3829 | 6886 |
| 2003 | 1626 | 1566 | 1910 | 0.0816279 | 0.0203524 | 0.10 | 0.05 | 2215 | 1698 | 846 | 0.610789331 | 0.146077718 | 0.10 | 0.05 | 3776 | 6758 |
| 2004 | 3574 | 2485 | 2262 | 0.0999628 | 0.0239871 | 0.10 | 0.05 | 5299 | 325 | 725 | 0.784045392 | 0.196083812 | 0.10 | 0.05 | 3719 | 6630 |
| 2005 | 3791 | 4392 | 3637 | 0.1095876 | 0.0273383 | 0.10 | 0.05 | 3839 | 101 | 1074 | 0.527434166 | 0.130273116 | 0.10 | 0.05 | 3665 | 6521 |
| 2006 | 3280 | 2604 | 2487 | 0.1106704 | 0.0272807 | 0.10 | 0.05 | 3627 | 286 | 776 | 0.498629615 | 0.126910467 | 0.10 | 0.05 | 3618 | 6445 |
| 2007 | 3648 | 5899 | 2530 | 0.1302676 | 0.0319828 | 0.10 | 0.05 | 3954 |  | 51 | 0.513346456 | 0.129373774 | 0.10 | 0.05 | 3580 | 6408 |
| 2008 | 3403 | 3325 | 1337 | 0.1234710 | 0.0306334 | 0.10 | 0.05 |  | 434 |  | 0.429647712 | 0.107854128 | 0.10 | 0.05 | 3554 | 6412 |
| 2009 | 2040 | 3024 | 1210 | 0.1202198 | 0.0298245 | 0.10 | 0.05 | 41 | 3321 | 683 | 0.396132057 | 0.099936755 | 0.10 | 0.05 | 3537 | 6448 |
| 2010 | 3177 | 4422 | 1930 | 0.1147915 | 0.0293706 | 0.10 | 0.05 | 31 |  | 738 | 0.387155067 | 0.099871349 | 0.10 | 0.05 | 3526 | 6502 |
| 2011 | 2418 | 4088 | 788 | 0.1307411 | 0.0330840 | 0.10 | 0.05 |  | 5196 | 741 | 0.296471008 | 0.076526048 | 0.10 | 0.05 | 3515 | 6554 |
| 2012 | 2511 | 4821 | 728 | 0.1109459 | 0.0296547 | 0.10 | 0.05 | 2 | 4577 | 617 | 0.314148489 | 0.082256468 | 0.10 | 0.05 | 3499 | 6583 |
| 2013 | 1531 | 3197 | 811 | 0.1150255 | 0.0300177 | 0.10 | 0.05 | 14 | 3429 | 839 | 0.286797524 | 0.073779157 | 0.10 | 0.05 | 3473 | 6572 |
| 2014 | 933 | 2683 | 720 | 0.1246719 | 0.0312521 | 0.10 | 0.05 | 857 | 2587 | 664 | 0.273761628 | 0.068716752 | 0.10 | 0.05 | 3435 | 6511 |
| 2015 | 826 | 2907 | 603 | 0.1073448 | 0.0287230 |  | 0.05 |  | 2894 | 439 | 0.227177038 | 0.062438232 | 0.10 | 0.05 | 3383 | 6403 |
| 2016 | 172 | 3903 | 2 | 0.1110073 | 0.0303749 |  |  | 00 | 3649 | 33 | 0.242659637 | 0.068229306 | 0.10 | 0.05 | 3322 | 6255 |
| 2017 | 368 | 3399 | 33 | 0.1172561 | 0.0320803 | 0.10 | 0.05 | 456 | 3851 | 79 | 0.244486978 | 0.070399207 | 0.10 | 0.05 | 3254 | 6086 |
| 2018 | 124 | 2433 | 62 | 0.1081235 | 0.0298193 | 0.10 | 0.05 | 110 | 2421 | 105 | 0.359871126 | 0.103203608 | 0.10 | 0.05 | 3184 | 5909 |

## Faroes

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## West Greenland

Annual input data for NEAC PFA run-reconstruction \& NCL models (uncertainty values define uniform distribution around estimates used in Monte Carlo simulation).


## Appendix 4: Input data for Atlantic salmon used to do the runreconstruction and estimates of returns and spawners by size group and age group for North America

Appendix 4.i. Input data for the fishery at West Greenland used in the run reconstruction model.

| Year | WGHarv | WGUnHa rv | WGMean Wt | WGSamp leNAC | WGSamp leNEAC | WGProp NACMin | WGProp NACMax | WGProp1 SWNAC | WGProp1 SWNEAC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 0 | 0 | 3.00 | 0 | 0 | 0.2 | 0.5 | 0.900 | 1.000 |
| 1971 | 2689 | 0 | 3.14 | 0 | 0 | 0.28 | 0.4 | 0.945 | 0.964 |
| 1972 | 2113 | 0 | 3.44 | 0 | 0 | 0.34 | 0.37 | 0.945 | 0.964 |
| 1973 | 2341 | 0 | 4.18 | 0 | 0 | 0.39 | 0.59 |  | 0.9 |
| 1974 | 1917 | 0 | 3.58 | 0 | 0 | 0.39 | 0.46 | 0.945 | 0.964 |
| 1975 | 2030 | 0 | 3.12 | 0 | 0 | 0.4 | 0.48 | 0.945 | 0.964 |
| 1976 | 1175 | 0 | 3.04 | 0 | 0 | 0.38 | 0.48 | 0.9 | 0.964 |
| 1977 | 1420 | 0 | 3.21 | 0 | 0 | 0.38 |  | 0.94 | 0.96 |
| 1978 | 984 | 0 | 3.35 | 0 | 0 | 0.47 | 0.57 |  |  |
| 1979 | 1395 | 0 | 3.34 | 0 | 0 | 0.48 | 0.52 | 0.945 | 0.964 |
| 1980 | 1194 | 0 | 3.22 | 0 | 0 | 0.45 | 2. 5 | 0.945 | 0.964 |
| 1981 | 1264 | 0 | 3.17 | 0 | 0 | . 58 |  |  | 0.964 |
| 1982 | 1077 | 0 | 3.11 | 0 |  | 0.6 |  | 0.94 | 0.964 |
| 1983 | 310 | 0 | 3.10 | 0 |  | 0.38 |  | 0.945 | 0.964 |
| 1984 | 297 | 0 | 3.11 | 0 |  | 0.47 | . 3 | 0.945 | 0.964 |
| 1985 | 864 | 0 | 2.87 | 0 |  | 0.46 | 0.53 | 0.925 | 0.950 |
| 1986 | 960 | 0 | 3.03 |  | 0 | 0.48 | 0.66 | 0.951 | 0.975 |
| 1987 | 966 | 0 | 3.16 |  | 0 | 0 0.54 | 0.63 | 0.963 | 0.980 |
| 1988 | 893 | 0 | 3.18 |  |  | 0.38 | 0.49 | 0.967 | 0.981 |
| 1989 | 337 | 0 | 2.87 | 0 |  | 0.52 | 0.6 | 0.923 | 0.955 |
| 1990 | 274 | 0 | 2.69 |  |  | 0.7 | 0.79 | 0.957 | 0.963 |
| 1991 | 472 | 0 | 2.65 |  | 0 | 0.61 | 0.69 | 0.956 | 0.934 |
| 1992 | 237 | 0 | 2.81 | 0 | 0 | 0.5 | 0.57 | 0.919 | 0.975 |
| 1993 | 0 |  | 2.73 |  | 0 | 0.5 | 0.76 | 0.946 | 0.961 |
| 1994 | 0 |  | 2.73 |  | 0 | 0.5 | 0.76 | 0.946 | 0.961 |
| 1995 |  |  |  |  | 0 | 0.65 | 0.72 | 0.968 | 0.973 |
| 1996 | 92 | 20 | 2.88 | $\bigcirc$ | 0 | 0.71 | 0.76 | 0.941 | 0.961 |
| 1997 | 58 |  | 2.71 | 0 | 0 | 0.75 | 0.84 | 0.982 | 0.993 |
| 1998 |  |  | 2.78 | 0 | 0 | 0.73 | 0.84 | 0.968 | 0.994 |
|  |  |  | 3.08 | 0 | 0 | 0.84 | 0.97 | 0.968 | 1.000 |
|  |  |  | 2.57 | 344 | 146 | 0 | 0 | 0.974 | 1.000 |
|  |  |  | 3.00 | 1 | 1 | 0.67 | 0.71 | 0.982 | 0.978 |
| 2002 |  |  | 2.90 | 338 | 163 | 0 | 0 | 0.973 | 1.000 |
|  |  | 10 | 3.04 | 1212 | 567 | 0 | 0 | 0.967 | 0.989 |
|  |  | 10 | 3.18 | 1192 | 447 | 0 | 0 | 0.970 | 0.970 |
| 2005 | 17.3 | 10 | 3.31 | 585 | 182 | 0 | 0 | 0.924 | 0.967 |
|  | 23 | 10 | 3.24 | 857 | 326 | 0 | 0 | 0.930 | 0.988 |
| 2007 | 24.8 | 10 | 2.98 | 917 | 206 | 0 | 0 | 0.965 | 0.956 |
| 2008 | 28.6 | 10 | 3.08 | 1593 | 260 | 0 | 0 | 0.974 | 0.988 |
| 2009 | 28 | 10 | 3.50 | 1483 | 138 | 0 | 0 | 0.934 | 0.894 |
| 2010 | 43.1 | 10 | 3.42 | 991 | 249 | 0 | 0 | 0.982 | 0.975 |
| 2011 | 27.4 | 10 | 3.40 | 888 | 72 | 0 | 0 | 0.939 | 0.831 |
| 2012 | 34.5 | 10 | 3.44 | 1121 | 252 | 0 | 0 | 0.932 | 0.980 |
| 2013 | 47.7 | 10 | 3.35 | 938 | 211 | 0 | 0 | 0.949 | 0.966 |
| 2014 | 70.482 | 10 | 3.32 | 660 | 260 | 0 | 0 | 0.913 | 0.961 |
| 2015 | 60.9 | 10 | 3.37 | 1337 | 337 | 0 | 0 | 0.970 | 0.982 |
| 2016 | 30.249 | 10 | 3.18 | 864 | 438 | 0 | 0 | 0.935 | 0.955 |
| 2017 | 28 | 10 | 3.46 | 734 | 252 | 0 | 0 | 0.925 | 0.931 |
| 2018 | 39 | 10 | 2.97 | 814 | 165 | 0 | 0 | 0.974 | 0.974 |

Appendix 4.ii. Input data for sea fisheries on large salmon and small salmon from Newfoundland and Labrador used in the run reconstruction model. Labrador represents harvests from Labrador in indigenous fisheries for food, social and ceremonial purposes and the resident food fishery beginning in 1998.


Appendix 4.iii. Input data for sea fisheries on large salmon and small salmon from St Pierre \& Miquelon used in the run-reconstruction model.


## Appendix 4.iv. Input data for large salmon for Labrador used in the run-reconstruction.

LB_SFA1_ LB_SFA2_ LB_SFA14


## Appendix 4.v. Input data for small salmon for Labrador used in the run-reconstruction.



Appendix 4.vi. Input data for returns of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.


Appendix 4.vi. (Continued). Input data for returns of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.


Appendix 4.vii. Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

|  | Salmon Fishing Area 3 |  |  |  | Salmon Fishing Area 4 |  |  |  | Salmon Fishing Area 5 |  |  |  | Salmon Fishing Area 6 |  |  | Salmon Fishing Area 7 |  | Salmon Fishing Area 8 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Small salmon\| | Large salmon |  | Small salmon |  | Large salmon |  | Small salmon |  | Large salmon |  | Small salmon |  | Large salmon | Small salmon | Large salmon | Small salmon |  | Large salmon |  |
| Year | Spawners |  | Spawners |  | Spawne |  | Spawners |  | Spawners |  | Spawners |  | Spawners |  | Spawners | Spawners |  | Spawners |  | Spawners |  |
|  | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | min $\quad$ max | min Max | Min Max | Min | Max | Min | Max |
|  | SFA3SSm_LI | SFA3SSm_HISF | FA3SLG_LD; | g-HIS | FAASSm_LDSFA | SSm_HISF | FAASLG_LD; | SLg_HII | FA5SSm_LDFA | Sm_H0; | FA5SLg_LD | Lg_H0F | A6SSm_LD:A | Sm_HD; | :FA6SLG_LI FA6SLg_H0F | A7SSm_LIT-A |  | ABSSm_LD=A |  | FA8SLG_LII |  |
| 1970 | 1829 | 4443 | 154 | 736 | 11314 | 27478 | 910 | 4512 | 5194 | 12614 | 404 | 2058 | 196 | 476 | ${ }^{14} \quad 76$ | 113 | $3-18$ | 43 | 105 | L |  |
| 1971 | 1731 | 4205 | 135 | 687 | 8827 | 21437 | 688 | 3499 | 3920 | 9520 | 293 | 1541 | 128 | 312 | $10 \quad 51$ | 227 | 8 38 | 58 | 142 | 0 | 15 |
| 1972 | 1162 | 2822 | 98 | 468 | 8036 | 19516 | 655 | 3214 | 4422 | 10738 | 354 | 1762 | 278 | 674 | $23 \quad 112$ | $142 \quad 346$ | $12 \quad 57$ | 65 | 159 | 6 |  |
| 1973 | 2772 | 6732 | 232 | 1115 | 15657 | 38023 | 1275 | 6259 | 4928 | 11968 | 405 | 1974 | 583 | 1417 | 49 | 306 | $26 \quad 123$ | 219 | 533 | 15 | 84 |
| 1974 | 1958 | 4754 | 318 | 641 | 12537 | 30447 | 1983 | 4049 | 3820 | 9276 | 608 | 1237 | 707 | 1717 | 115 |  | 100 | 119 | 289 | 20 |  |
| 1975 | 2583 | 6273 | 520 | 1041 | 13867 | 33677 | 2628 | 5421 | 4639 | 11265 | 912 | 1846 | 219 | 533 | 43 |  | 38 | 203 | 493 | 41 | 82 |
| 1976 | 2210 | 5366 | 379 | 759 | 15594 | 37870 | 2495 | 5177 | 4429 | 10755 | 697 | 1459 | 576 | 1400 | 97196 | 170 | 24 | 187 | 453 | 32 | 64 |
| 1977 | 3570 | 8670 | 478 | 960 | 19591 | 47577 | 1559 | 4204 | 10771 | 26157 | 1410 | 2864 | 936 | 2272 |  |  | 48 | 189 | 459 | 26 | 51 |
| 1978 | 1769 | 4295 | 149 | 298 | 20473 | 49719 | 1229 | 2959 | 6669 | 16195 | 536 | 1100 | 691 | 1677 |  | 561 | 1938 | 103 | 249 | 9 | 17 |
| 1979 | 4760 | 11560 | 390 | 779 | 18727 | 45481 | 1206 | 2740 | 3106 | 7542 | 234 | 489 | 569 | 1383 |  |  | $24 \quad 48$ | 233 | 567 | 19 |  |
| 1980 | 4067 | 9877 | 224 | 485 | 21966 | 53346 | 903 | 2312 | 6305 | 15311 | 376 | 780 | 747 | 1813 |  | 238 | $14 \quad 30$ | 280 | 680 | 18 | 36 |
| 1981 | 5502 | 13362 | 1042 | 2087 | 31584 | 76704 | 5637 | 11635 | 8139 | 19765 | 1511 | 3056 | 1412 |  |  |  | $53 \quad 107$ | 180 | 436 | 34 | 68 |
| 1982 | 6146 | 14926 | 124 | 336 | 23270 | 56514 | 544 | 1346 | 5677 | 13787 | 143 | 338 | 672 |  |  |  | 15 | 198 | 482 | 0 |  |
| 1983 | 3773 | 9163 | 245 | 493 | 20893 | 50739 | 1073 | 2443 | 5500 | 13356 | 191 | 551 | 691 |  | $35 \quad 81$ | 788 | 9 | 96 | 232 | 1 | 8 |
| 1984 | 2531 | 6525 | 55 | 540 | 25033 | 64536 | 533 | 5322 | 6835 | 17620 | 149 | 1456 | 789 |  | 12 | 626 | 48 | 200 | 515 | 4 |  |
| 1985 | 3462 | 8569 | 72 | 683 | 32218 | 79741 | 671 | 6352 | 9208 | 22791 | 192 | 1816 | 1134 | 806 | 24 | $296 \quad 733$ | 58 | 272 | 674 | 6 | 54 |
| 1986 | 2054 | 5126 | 70 | 413 | 24722 | 6170 | 840 | 4977 | 10782 | 26910 | 366 | 2170 | 1184 |  | 40 | 271677 | 95 | 367 | 916 | 12 |  |
| 1987 | 1655 | 3895 | 57 | 318 | 16032 | 37722 | 556 | 3078 | 4892 | 11511 | 170 | 939 | 403 |  | 14 | $82 \quad 194$ | 16 | 126 | 297 | 4 | 24 |
| 1988 | 4868 | 11888 | 159 | 956 | 27317 | 66712 | 892 | 5367 | 11549 | 28204 | 377 | 2269 | 1189 |  | 39 | $355 \quad 867$ | $12 \quad 70$ | 219 | 535 | 7 | 43 |
| 1989 | 2266 | 5376 | 90 | 461 | 11623 | 27581 | 461 | 2365 | 4350 | 10323 | 172 |  |  |  |  | 203481 | 41 | 304 | 721 | 12 |  |
| 1990 | 5032 | 10098 | 236 | 920 | 16583 | 33273 | 776 | 3033 | 7071 | 14188 | 331 |  | 978 |  |  | 144288 | $7 \quad 26$ | 252 | 505 | 12 | 46 |
| 1991 | 4334 | 7965 | 193 | 750 | 16113 | 29607 | 718 | 2788 | 6745 | 12395 | 301 |  |  | 1126 | 106 | $119 \quad 218$ | $5 \quad 21$ | 36 | 67 | 2 |  |
| 1992 | 9844 | 21262 | 415 | 4094 | 33228 | 71898 | 1407 | 13866 | 12175 | 26363 | 516 |  |  |  | $61 \quad 603$ | 252545 | $11 \quad 105$ | 0 | 0 | 0 |  |
| 1993 | 10054 | 20961 | 400 | 1599 | 39162 | 81344 | 1590 | 6226 | 14370 | 29779 | 576 |  |  |  | $90 \quad 351$ | $404 \quad 831$ | 16 63 | 369 | 760 | 15 | 58 |
| 1994 | 9146 | 24802 | 749 | 3247 | 20576 | 55760 | 1644 | 7259 | 6855 | 18510 | 560 |  |  |  | $30 \quad 133$ | $46 \quad 122$ | $4 \quad 16$ | 79 | 212 | 6 |  |
| 1995 | 7409 | 21791 | 580 | 2636 | 22872 | 67179 | 1801 | 8135 | 8122 | 23776 | 642 | 2880 |  | 831 | $23 \quad 100$ | 173501 | $14 \quad 60$ | 135 | 397 | 11 | 48 |
| 1996 | 15729 | 39860 | 1412 | 4247 | 42346 | 107268 | 3757 | 11383 | 14095 | 35586 | 1263 | 3787 |  | 1317 | $46 \quad 139$ | $124 \quad 311$ | $11 \quad 33$ | 180 | 457 | 16 |  |
| 1997 | 9422 | 19095 | 1209 | 3954 | 19309 | 39107 | 2467 | 8083 | 5228 | 10547 | 668 | 2177 |  | 384 | $24 \quad 79$ | $49 \quad 99$ | 20 | 48 | 98 | 6 | 20 |
| 1998 | 16390 | 24345 | 1933 | 6969 | 43559 | 64785 | 5160 | 18599 | 9943 |  | 1155 | 4201 |  | 673 | 53191 | $212 \quad 313$ | $25 \quad 88$ | 135 | 201 | 16 | 57 |
| 1999 | 11804 | 18173 | 1279 | 4189 | 52390 | 80698 | 5650 | 18583 | 8832 |  |  | 3126 |  | 528 | $37 \quad 121$ | $58 \quad 90$ | $6 \quad 21$ | 119 | 188 | 14 |  |
| 2000 | 17003 | 23723 | 1449 | 3711 | 32879 | 45946 | 2803 | 7201 | 10897 |  |  | 2377 | 938 | 1386 | $84 \quad 217$ | $140 \quad 195$ | $12 \quad 31$ | 88 | 125 | 8 | 20 |
| 2001 | 9861 | 13889 | 892 | 2089 | 33365 | 45682 | 3023 | 7086 |  |  |  |  |  | 342 | $23 \quad 53$ | $42 \quad 59$ | $4 \quad 9$ | 17 | 23 | 2 |  |
| 2002 | 8620 | 13856 | 671 | 1994 | 28099 | 45208 | 2175 | 6498 |  |  |  |  |  | 319 | $15 \quad 45$ | $0 \quad 0$ | $0 \quad 0$ | 55 | 91 | 5 | 14 |
| 2003 | 19386 | 23938 | 1085 | 3478 | 67026 | 82739 | 3738 | 12001 |  |  |  |  | 412 | 508 | $23 \quad 74$ | $94 \quad 116$ | $5 \quad 17$ | 47 | 58 | 3 |  |
| 2004 | 6942 | 11402 | 390 | 1680 | 43104 | 70785 | 2430 |  |  | 119 |  |  | 158 | 259 | $9 \quad 38$ | $0 \quad 0$ | $0 \quad 0$ | 35 | 58 | 2 |  |
| 2005 | 5056 | 17534 | 473 | 2664 | 28896 | 100323 | 2695 |  |  |  | 394 | 2205 | 92 | 314 | 47 | $0 \quad 0$ | 00 | 18 | 69 | 2 | 11 |
| 2006 | 9402 | 15839 | 1228 | 3216 | 37156 | 62732 | 4925 |  |  |  | 969 | 2554 | 61 | 102 | 20 | $0 \quad 0$ | $0 \quad 0$ | 141 | 244 | 20 |  |
| 2007 | 9147 | 19842 | 1171 | 3818 | 32243 | 70143 | 4122 |  |  |  | 978 | 3196 | 68 | 148 | 28 | 112245 | $12 \quad 45$ | 15 | 33 | 2 | 6 |
| 2008 | 11799 | 21183 | 1045 | 3379 | 53591 | 96443 |  |  |  | 17405 | 867 | 2791 | 274 | 497 | $22 \quad 78$ | $69 \quad 125$ | $4{ }^{18}$ | 159 | 292 | 15 |  |
| 2009 | 11205 | 22795 | 779 | 5080 | 49881 | 101728 |  | 22732 |  | 18065 | 622 | 4049 | 412 | 834 | $28 \quad 185$ | $0 \quad 0$ | $0 \quad 0$ | 111 | 228 | 7 | 51 |
| 2010 | 18364 | 23569 | 1595 | 4581 | 69075 | 88772 |  |  | 20114 | 5822 | 1754 | 5028 | 874 | 1120 | $76 \quad 217$ | $183 \quad 235$ | $16 \quad 46$ | 93 | 120 | 8 | 24 |
| 2011 | 13193 | 24264 | 1291 | 6261 | 50806 | 93428 |  |  |  |  | 1176 | 5734 | 716 | 1314 | $70 \quad 339$ | $83 \quad 153$ | $8 \quad 39$ | 220 | 412 | 22 | 108 |
| 2012 | 21149 | 31048 | 1639 | 4394 | 64959 | 95436 |  |  | 14554 | 3 | 1140 | 3039 | 738 | 1086 | $57 \quad 154$ | $100 \quad 147$ | 21 | 361 | 533 | 25 | 73 |
| 2013 | 7822 | 12219 | 495 | 4039 | 44838 | 70136 |  |  |  | 23531 | 976 | 7833 | 721 | 1128 | $44 \quad 373$ | $241 \quad 379$ | $16 \quad 127$ | 102 | 162 | 7 | 55 |
| 2014 | 11917 | 19224 | 1112 | 2885 | 35599 |  |  |  |  | 11822 | 680 | 1773 | 452 | 728 | $39 \quad 106$ | $106 \quad 176$ | $8 \quad 25$ | 87 | 142 | 8 |  |
| 2015 | 17382 | 29684 | 1679 | 4862 | 69539 |  |  |  |  | 24765 | 1399 | 4055 | 842 | 1435 | $81 \quad 234$ | $117 \quad 203$ | $12 \quad 34$ | 280 | 484 | 28 | 81 |
| 2016 | 10280 | 22007 | 1078 | 3438 | 32370 |  |  |  |  | 14754 | 733 | 2317 | 373 | 799 | $41 \quad 127$ | $92 \quad 202$ | $10 \quad 32$ | 174 | 375 | 19 | 59 |
| 2017 | 6134 | ${ }^{18793}$ | 555 | 2476 |  |  |  |  |  | 15190 | ${ }^{466}$ | 2024 | 85 |  | 38 | $46 \quad 147$ | 17 | 89 | 279 | 9 |  |
| 2018 | 31981 | 59717 | 1412 | 6341 |  |  |  |  |  | 25340 | 592 | 2688 | - | , | 0 | 446834 | $20 \quad 89$ | 221 | 415 | 10 | 44 |

Appendix 4.vii. (Continued). Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

|  | Salmon Fishing Area 9 |  |  |  | Salmon Fishing Area 10 |  |  |  | Salmon Fishing Area 11 |  |  |  | Salmon Fishing Area 12 |  |  |  | Salmon Fishing Area 13 |  |  |  | Salmon Fishing Area 14A |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Small salmon |  | Large salmon |  | Small salmon |  | Large salmon |  | Small salmon |  | Large salmon |  | Small salmon |  | Large salmon |  | Small salmon |  | Large salmon |  | Small salmon |  | Large salmon |  |
| Year | Spawners |  | Spawners |  | $\begin{array}{\|l\|l} \hline \text { Spawners } \\ \text { Min } & \text { Max } \\ \hline \end{array}$ |  | Spawners |  | Spawners |  | $\begin{aligned} & \text { Spawners } \\ & \text { Min } \\ & \hline \end{aligned}$ |  | Spawners Min | Max | $\begin{array}{\|l} \begin{array}{l} \text { Spawners } \\ \text { Min } \end{array} \\ \hline \text { mSEA12SLa } \end{array}$MFA12SL | $\frac{\text { Max }}{\text { SFA12SLK }}$ | $\begin{aligned} & \text { Spawners } \\ & \text { Min } \end{aligned}$ |  | Spawners Min | Max | $\begin{array}{\|l} \hline \begin{array}{l} \text { Spawners } \\ \text { Min } \end{array} \\ \hline \end{array}$ | Max | ( ${ }_{\text {Spawners }}$ | Max |
|  | Min | Max | Min | Max |  |  | Min | Max | Min | Max |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | SFA9SSm | SFA9SSm | SFA9SLI | ISFA9SLg | SFA10SSm | SFA10SSn | SFA10SLg | SFA10SLg | 11 SSm | SFA |  | SFA11SLg |  | SFA1 |  |  | A13SSm | FA13SS | SFA13SLg | SFA13SL9 | SFA14ASSm | SFA14ASSm | -SFA | ISFA14ASLg H |
| 1970 | 4417 | 10727 | 361 | 1768 | 1402 | 3406 | 112 | 558 | 11732 | 28492 | 918 | 4653 | 1748 | 4244 | 69 | 625 | 16203 | 28543 | 1608 | 3417 | 10372 | 25188 | 134 | 2340 |
| 1971 | 3780 | 9180 | 301 | 1504 | 2165 | 5259 | 166 | 855 | 9473 | 23007 | 736 | 3752 | 1059 | 2573 | 74 |  | 489 | 30629 | 1633 | 3705 | 8766 | 21290 | 0 | 1850 |
| 1972 | 2658 | 6454 | 217 | 1063 | 1323 | 3213 | 108 | 529 | 11445 | 27795 | 882 | 4525 | 2165 | 5259 | 163 | 52 |  | 29188 | 2004 | 4066 | 5640 | 13696 | 83 | 1283 |
| 1973 | 5040 | 12240 | 406 | 2011 | 4165 | 10115 | 310 | 1636 | 11331 | 27517 | 923 | 4530 | 1507 | 3661 | 102 | 582 |  | 29 | 1911 | 3808 | 12325 | 29931 | 91 | 2713 |
| 1974 | 3486 | 8466 | 565 | 1140 | 2828 | 6868 | 452 | 918 | 10444 | 25364 | 1682 | 3403 | 1535 | 3729 | 24 | 493 |  |  | 1997 | 3341 | 7280 | 17680 | 789 | 1692 |
| 1975 | 4368 | 10608 | 874 | 1754 | 996 | 2420 | 192 | 392 | 10502 | 25506 | 2076 | 4192 | 1190 | 2890 |  | 459 | 21668 | 421 | 3611 | 6538 | 11242 | 27302 | 417 | 925 |
| 1976 | 3787 | 9197 | 639 | 1291 | 1703 | 4137 | 283 | 576 | 9716 | 23596 | 1629 | 3301 | 693 | 1683 |  |  | 18999 | 36519 | 2752 | 4862 | 17222 | 41826 | 1337 | 2774 |
| 1977 | 2520 | 6120 | 331 | 671 | 2560 | 6216 | 341 | 686 | 9557 | 23211 | 1272 | 2563 | 1302 | 3162 |  |  | 10898 | 18528 | 1828 | 2549 | 13316 | 32340 | 194 | 859 |
| 1978 | 3040 | 7384 | 240 | 497 | 3722 | 9038 | 273 | 587 | 9324 | 22644 | 770 | 1558 | 854 | 2074 |  | 24 |  | 22392 | 3861 | 4434 | 7562 | 18366 | 194 | 460 |
| 1979 | 3976 | 9656 | 311 | 636 | 1981 | 4811 | 154 | 316 | 8003 | 19437 | 648 | 1304 | 1710 | 4154 | , | 70 |  | 25820 | 1070 | 1749 | 15349 | 37275 | 174 | 408 |
| 1980 | 5551 | 13481 | 295 | 651 | 3556 | 8636 | 201 | 429 | 11828 | 28724 | 715 | 1474 | 1913 |  | 94 | 17 | 625 | 30958 | 4243 | 4920 | 8734 | 21210 | 514 | 1208 |
| 1981 | 4345 | 10551 | 773 | 1598 | 3073 | 7463 | 555 | 1138 | 16478 | 40018 | 3088 | 6217 | 2473 | 6007 | 453 |  | 22059 | 36689 | 4485 | 6789 | 13725 | 33331 | 953 | 2043 |
| 1982 | 4258 | 10342 | 114 | 260 | 2931 | 7117 | 91 | 192 | 17122 | 41582 | 537 | 1127 | 3628 | 8812 | 110 |  | 22062 | 37132 | 2847 | 3236 | 11114 | 26990 | 2987 | 6082 |
| 1983 | 5374 | 13050 | 281 | 634 | 2660 | 6460 | 95 | 270 | 11128 | 27024 | 703 | 1433 | 1556 | 3780 | 94 | 196 | 14491 | 25364 | 3855 | 4490 | 8867 | 21533 | 1635 | 3338 |
| 1984 | 5725 | 14759 | 120 | 1216 | 3684 | 9498 | 79 | 783 | 17748 | 45755 | 374 | 3769 | 4860 | 12529 | 38 | 968 | 18413 | 30081 | 1987 | 3401 | 12155 | 31336 | 179 | 2504 |
| 1985 | 4625 | 11448 | 96 | 912 | 3505 | 8674 | 73 | 691 | 15390 | 38091 | 320 | 3034 | 2899 | 7176 | 57 | 569 | 10726 | 20203 | 1349 | 2482 | 9583 | 23719 | 197 | 1887 |
| 1986 | 6113 | 15257 | 208 | 1231 | 4084 | 10192 | 139 | 822 | 14754 | 36822 | 501 | 2970 | 2495 | 6228 | 81 | 499 | 15535 | 29570 | 2013 | 3583 | 13381 | 33396 | 445 | 2683 |
| 1987 | 2549 | 5998 | 88 | 489 | 1261 | 2968 | 44 | 242 | 11258 | 26488 | 391 | 2162 | 2443 | 5749 |  | 466 | 13611 | 26307 | 1512 | 2988 | 13583 | 31960 | 467 | 2604 |
| 1988 | 3806 | 9295 | 124 | 748 | 3166 | 7731 | 103 | 622 | 13952 | 34073 | 456 |  | 3917 |  | 126 | 767 | 17945 | 35263 | 1909 | 3877 | 17329 | 42319 | 549 | 3388 |
| 1989 | 4037 | 9580 | 160 | 821 | 2757 | 6542 | 109 | 561 | 9087 | 21564 | 360 |  | 719 | 4080 | 67 | 349 | 6980 | 12982 | 836 | 1552 | 9833 | 23334 | 385 | 1996 |
| 1990 | 5466 | 10968 | 256 | 1000 | 2446 | 4908 | 115 | 447 | 13024 | 26132 | 610 | 238 | 2507 | 5031 | 114 | 456 | 14866 | 24531 | 1744 | 3051 | 14793 | 29682 | 679 | 2692 |
| 1991 | 1844 | 3389 | 82 | 319 | 758 | 1392 | 34 | 131 | 6103 | 11215 | 272 | 1056 |  |  | 93 | 365 | 11037 | 15757 | 1689 | 2413 | 11742 | 21576 | 512 | 2020 |
| 1992 | 4334 | 9378 | 183 | 1809 | 1496 | 3287 | 65 | 642 | 14239 | 30854 | 605 | 5958 |  | 8657 | 162 | 1667 | 20506 | 38635 | 2992 | 8833 | 30096 | 65023 | 1234 | 12488 |
| 1993 | 9956 | 20502 | 400 | 1559 | 4809 | 9967 | 194 | 761 | 21423 | 44150 | 861 | 3359 | 5176 | 10666 | 207 | 810 | 22341 | 41708 | 2544 | 4673 | 27010 | 55787 | 1058 | 4221 |
| 1994 | 2124 | 5723 | 172 | 746 | 1804 | 4848 | 144 | 630 | 5295 | 14449 | 430 | 1891 | 1949 | 5253 | 154 | 681 | 15381 | 30444 | 3207 | 5611 | 9385 | 25327 | 742 | 3286 |
| 1995 | 3887 | 11386 | 304 | 1376 | 3218 | 9378 | 253 | 1133 | 7770 | 22930 | 625 | 2792 | 1689 | 4922 | 130 | 592 | 20570 | 43329 | 1607 | 4860 | 15218 | 44587 | 1187 | 5385 |
| 1996 | 4868 | 12304 | 431 | 1304 | 6687 | 16880 | 592 | 1789 | 15226 | 38638 | 1362 | 4113 | 4082 | 10295 | 358 | 1088 | 29056 | 60152 | 3199 | 6852 | 26584 | 67085 | 2357 | 7115 |
| 1997 | 2927 | 5918 | 372 | 1221 | 4086 | 8257 | 519 | 1702 | 13304 | 26995 | 1718 | 5602 | 3655 | 7401 | 464 | 1527 | 25508 | 40599 | 3985 | 8266 | 20359 | 41117 | 2578 | 8467 |
| 1998 | 3937 | 5840 | 458 | 1663 | 6606 | 9777 | 771 | 2779 | 7024 | 10457 | 36 | 3009 | 1968 | 2925 | 225 | 831 | 18279 | 24417 | 3031 | 6918 | 19992 | 29721 | 2347 | 8507 |
| 1999 | 3401 | 5230 | 359 | 1195 | 4313 | 6642 | 455 | 1520 | 8086 |  |  | 2 | 958 | 1477 | 102 | 339 | 28647 | 37098 | 3760 | 7621 | 22659 | 34941 | 2402 | 8013 |
| 2000 | 6913 | 9645 | 581 | 1501 | 6664 | 9322 | 534 | 1429 | 14895 | 20901 | 1288 |  | 2291 | 3208 | 195 | 504 | 48055 | 61508 | 5250 | 9779 | 32314 | 45127 | 2731 | 7044 |
| 2001 | 1709 | 2339 | 151 | 359 | 2436 | 3338 | 215 | 513 | 04 | 1070 | 714 | 1671 | 1818 | 2497 | 162 | 386 | 31037 | 39405 | 3536 | 6297 | 17331 | 23744 | 1559 | 3674 |
| 2002 | 1562 | 2518 | 118 | 360 | 3049 | 4901 | 231 | 699 | 3347 | 118 | 581 | 1716 | 1896 | 3053 | 147 | 439 | 28083 | 40025 | 3313 | 6330 | 21764 | 35007 | 1668 | 5013 |
| 2003 | 1985 | 2454 | 109 | 355 | 3368 | 4162 | 185 | 603 | 2701 | 693 | 703 | 2276 | 5282 | 6528 | 288 | 943 | 45027 | 52657 | 4206 | 8218 | 36597 | 45189 | 1988 | 6506 |
| 2004 | 1674 | 2749 | 91 | 402 | 3210 | 5273 | 177 | 744 |  | 9544 | 660 | 2882 | 2704 | 4452 | 149 | 655 | 43889 | 60513 | 4074 | 8883 | 26116 | 42892 | 1429 | 6282 |
| 2005 | 1478 | 5264 | 130 | 794 | 2171 | 7572 | 194 | 142 |  | 16992 | 456 | 2591 | 2062 | 7282 | 191 | 1107 | 33349 | 81031 | 4320 | 12691 | 13676 | 47376 | 1246 | 7163 |
| 2006 | 3791 | 6397 | 498 | 1302 | 4627 | 7824 | 602 | 1590 | 9554 | 16155 | 1271 | 3310 | 2986 | 5056 | 392 | 1032 | 46296 | 67532 | 8247 | 14807 | 24532 | 41334 | 3210 | 8400 |
| 2007 | 2063 | 4502 | 263 | 867 | 3047 | 6636 | 387 | 1275 |  | 706 | 636 | 2071 | 2442 | 5323 | 314 | 1027 | 29402 | 54734 | 4511 | 10780 | 17446 | 37934 | 2222 | 7293 |
| 2008 | 3285 | 5948 | 293 | 955 | 3971 | 7202 |  | 1154 | 9314 | 5832 | 841 | 2711 | 2178 | 3940 | 193 | 631 | 43277 | 66465 | 3580 | 9346 | 21887 | 39305 | 1915 | 6246 |
| 2009 | 2835 | 5834 | 198 | 1311 | 4193 | 8665 |  | 1957 | 6203 | 12763 | 442 | 2877 | 1450 | 2970 | 100 | 664 | 31106 | 50196 | 3526 | 10610 | 17820 | 36229 | 1200 | 8032 |
| 2010 | 5703 | 7334 | 496 | 1432 | 7062 | 9081 |  |  | 6859 | 8842 | 604 | 1742 | 2606 | 3347 | 226 | 651 | 49703 | 58889 | 5478 | 10747 | 27468 | 35298 | 2358 | 6848 |
| 2011 | 4364 | 8077 | 433 | 2099 | 7477 | 13826 | 716 | 66 | 5696 | 10554 | 564 | 2744 | 2074 | 3827 | 203 | 990 | 33849 | 62267 | 3160 | 15915 | 20249 | 37231 | 1953 | 9575 |
| 2012 | 5898 | 8720 | 471 | 1256 | 7488 | 11027 |  |  |  | 8819 | 468 | 1255 | 2348 | 3450 | 184 | 490 | 44778 | 65820 | 3395 | 9251 | 31467 | 46268 | 2451 | 6571 |
| 2013 | 3973 | 6228 | 254 | 2071 | 6681 |  |  |  | 30 | 9621 | 398 | 3213 | 1701 | 2668 | 104 | 885 | 28314 | 44299 | 3301 | 13638 | 11746 | 18341 | 734 | 6050 |
| 2014 | 2222 | 3597 | 205 | 538 | 3340 | 417 | 03 |  | 2701 | 4382 | 255 | 663 | 1374 | 2218 | 125 | 330 | 27090 | 43731 | 3558 | 8032 | 19410 | 31358 | 1782 | 4682 |
| 2015 | 3897 | 6679 | 375 | 1095 | 4182 | 7159 | 393 | 1164 | 4776 | 8201 | 465 | 1351 | 1622 | 2778 | 145 | 444 | 41399 | 70742 | 5248 | 12970 | 27736 | 47369 | 2666 | 7747 |
| 2016 | 2666 | 5928 | 301 | 957 | 3783 | 8189 | 05 | 1293 | 3906 | 8514 | 419 | 1347 | 1612 | 3499 | 163 | 543 | 22477 | 48428 | 3316 | 9211 | 17030 | 36504 | 1792 | 5712 |
| 2017 | 3517 | 10973 | 321 | 1452 | 3454 | 0695 | 317 | 1416 | 2040 | 6294 | 194 | 840 | 2074 | 6403 | 183 | 840 | 26209 | 80920 | 3148 | 12915 | 25140 | 77282 | 2265 | 10177 |
| 2018 | 3878 | 7275 | 169 | 773 | 8798 |  | 377 | 1740 | 2892 | 5415 | 129 | 577 | 3646 | 6849 | 149 | 718 | 32388 | 60745 | 3474 | 9684 | 16924 | 31782 | 716 | 3356 |

## Appendix 4.viii. Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.



Appendix 4.viii. (Continued). Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.


## Appendix 4.ix. Input data for small salmon and large salmon returns to Québec by category of data used in the run-reconstruction.

| Year | $\underset{L}{\text { QCSmC1_ }}$ | $\begin{aligned} & \text { QCSmC2_ } \\ & \mathrm{L} \end{aligned}$ | $\mathrm{QCSmCl}_{\mathrm{L}}^{\mathrm{C}} \mathrm{~L}$ |  | ${ }_{L}^{\text {QCSmC5_ }}$ | $\begin{aligned} & \text { QCSmC6_ } \\ & \text { L } \\ & \text { L } \end{aligned}$ | $\mathrm{QCSmFn}_{\mathrm{L}}^{\mathrm{Q}} \mathrm{O}_{\mathrm{H}}^{\mathrm{Q}}$ | $\begin{aligned} & \text { QCSmC1_ } \\ & \mathrm{H} \end{aligned}$ | $\begin{aligned} & \text { QCSmC2_ } \\ & \text { H } \end{aligned}$ | $\begin{aligned} & \text { QCSmC3_C } \\ & \mathrm{H} \end{aligned}$ | $\begin{aligned} & \text { QCSmC4_O } \\ & \text { H } \end{aligned}$ | CSmC5_ | $\begin{array}{ll} \text { QCSmC6_ QC } \\ \text { H } \end{array}$ | SmFn_ | QCLEC1_L | QCLgC2_L | QCLgC3_L | QCLıC4_L | C5_L | LgC6_L | gFn_L | CLgC1_ | $\begin{aligned} & \text { QCLgC2_ } \\ & H \end{aligned}$ | $\begin{array}{ll} \mathrm{QCLgC3} & \mathrm{O} \\ \mathrm{H} & \mathrm{H} \end{array}$ | $\begin{array}{ll} \text { QCLgC4_ } \\ \mathrm{H} & \mathrm{O} \\ \mathrm{H} \end{array}$ | $\begin{aligned} & \text { QCLgC5_ } \\ & \text { H } \end{aligned}$ | $\begin{aligned} & \text { QCLLC6_ } \\ & \text { H } \end{aligned}$ | $\begin{aligned} & \text { QCLLgFn_ } \\ & \text { H } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  | 0 | 0 | 0 | 0 | 0 |  | 0 |  |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 3830 | 5434 | 2955 | 460 | 1670 | 5160 | 267 | 4085 | 5639 | 6053 | 792 | 2784 | 8599 | 445 | 14119 | 9501 | 22 | 407 | 3712 | 5071 | 329 | 15631 | 9788 | 6035 | 6477 | 6187 | 8452 | 548 |
| 1985 | 5266 | 2271 | 1767 | 210 | 5449 | 4384 | 267 | 5869 | 2336 | 3586 | 352 | 9224 | 7307 | 445 | 14015 | 7028 | 836 | 345 | 9215 | 3351 | 329 | 15611 | 7281 | 7809 | 577 | 15827 | 5586 | 548 |
| 1986 | 8648 | 5193 | 2396 | 63 | 6719 | 5133 | 267 | 9471 | 5321 | 4895 | 107 | 11198 | 8555 |  | 589 | 8598 | 152 | 35 | 5877 | 4971 | 329 | 20602 | 8839 | 12596 | 61 | 9795 | 8284 | 548 |
| 1987 | 10043 | 4775 | 3852 | 327 | 8396 | 5501 | 267 | 10869 | 4910 | 7875 | 546 | 13993 | 9168 |  |  | 6715 | 5178 | 273 | 6335 | 3012 | 329 | 19017 | 6889 | 10575 | 458 | 10558 | 5019 | 548 |
| 1988 | 11190 | 5968 | 4404 | 468 | 8440 | 6423 | 267 | 12244 | 6133 | 8962 | 780 | 14067 | $10 \%$ |  |  |  | 7540 | 346 | 6789 | 4781 | 329 | 22979 | 6618 | 15336 | 576 | 11315 | 7969 | 548 |
| 1989 | 10121 | 4743 | 2924 | 301 | 6744 | 5622 | 267 | 10910 | 4878 | 5940 | 503 | 11240 |  |  | 20278 | 8503 | 5530 | 278 | 5718 | 4567 | 329 | 21906 | 8736 | 11252 | 465 | 9531 | 7611 | 548 |
| 1990 | 12245 | 7332 | 4377 | 694 | 7096 | 2976 | 377 | 13278 | 7511 | 8917 | 1158 | 11826 |  |  | 17098 | 10803 | 8164 | 1365 | 5179 | 2424 | 442 | 18222 | 11041 | 16613 | 2276 | 8631 | 4040 | 737 |
| 1991 | 9554 | 5851 | 3776 | 349 | 5009 | 2001 | 256 | 10249 | 5987 | 7679 | 584 | 8348 |  |  | 191 | 6988 | 7183 | 696 | 3856 | 357 | 242 | 20443 | 7192 | 14602 | 1161 | 6427 | 595 | 403 |
| 1992 | 9188 | 6928 | 4567 | 428 | 5131 | 3462 | 243 | 9847 | 7144 | 9297 | 715 | 8552 |  |  | 18392 | 7360 | 7930 | 372 | 2687 | 1503 | 461 | 19578 | 7560 | 16149 | 622 | 4478 | 2505 | 769 |
| 1993 | 8143 | 6325 | 3973 | 1029 | 4315 | 1447 | 525 | 8883 | 6517 | 8075 | 1717 | 7192 | 2412 |  | 14578 | 10133 | 2866 | 373 | 2649 | 333 | 423 | 15454 | 11463 | 5849 | 624 | 4414 | 555 | 705 |
| 1994 | 8707 | 5928 | 3840 | 1051 | 4011 | 437 | 408 | 9442 | 6129 | 7828 | 1753 | 6686 | 729 |  | 16538 | 9172 | 2644 | 506 | 2853 | 145 | 427 | 17594 | 10241 | 5411 | 845 | 4755 | 242 | 712 |
| 1995 | 6943 | 3439 | 2697 | 1017 | 3853 | 434 | 184 | 7538 | 3527 | 5471 |  | 6422 |  |  | 21658 | 9598 | 1926 | 813 | 4390 | 154 | 246 | 22968 | 10936 | 3915 | 1358 | 7317 | 256 | 410 |
| 1996 | 15010 | 1809 | 3600 | 477 | 4666 | 500 | 120 | 16122 | 1923 | 7370 |  | 7816 |  |  | 22679 | 5822 | 3843 | 577 | 2486 | 135 | 113 | 24117 | 6941 | 7844 | 964 | 4155 | 225 | 189 |
| 1997 | 11491 | 201 | 3457 | 292 | 3529 | 462 | 58 | 12089 | 242 |  | - 487 | 882 |  |  | 18106 | 4221 | 2816 | 333 | 2865 | 138 | 48 | 19154 | 5154 | 5768 | 553 | 4775 | 229 | 80 |
| 1998 | 11285 | 1183 | 3578 | 328 | 5121 | 1127 | 58 | 11849 | 1406 |  |  |  |  | 97 | 13180 | 4927 | 2861 | 347 | 2790 | 291 | 48 | 13891 | 5962 | 5907 | 592 | 4649 | 485 | 80 |
| 1999 | 10877 | 708 | 3194 | 1868 | 5401 | 1429 | 0 | 11556 |  | 6536 |  |  |  |  | 16912 | 842 | 2554 | 3661 | 3870 | 492 | 0 | 17700 | 995 | 5232 | 6103 | 6450 | 838 | 0 |
| 2000 | 11886 | 429 | 1116 | 602 | 7399 | 633 | 0 | 12635 |  | 2284 |  | 14050 | 1055 |  | 14568 | 619 | 3901 | 560 | 6420 | 563 | 0 | 15300 | 669 | 7947 | 933 | 10700 | 949 | 0 |
| 2001 | 8050 | 185 | 2632 | 266 | 3225 | 728 | 0 | 8588 |  | 5392 |  | 5374 | 1213 |  | 17837 | 633 | 5320 | 241 | 3988 | 556 | 0 | 18889 | 879 | 10914 | 402 | 6647 | 926 | 0 |
| 2002 | 14599 | 31 | 3189 | 689 | 4333 | 1448 | 0 |  |  |  | 149 | 7222 | 2414 |  | 12335 | 8 | 4515 | 339 | 2103 | 345 | - | 13001 | 9 | 9277 | 565 | 3505 | 575 |  |
| 2003 | 11362 | 0 | 3203 | 721 | 3566 | 1512 | 0 |  |  |  | 1201 | 5944 | 2520 | 0 | 21853 | 0 | 5787 | 269 | 4889 | 384 | 0 | 22893 | 0 | 11779 | 449 | 8148 | 641 | 0 |
| 2004 | 13747 | 107 | 6526 | 284 | 4889 | 1639 | - |  | 127 |  | 474 | 8149 | 2731 | 0 | 18369 | 107 | 4870 | 357 | 4432 | 401 | 0 | 19043 | 126 | 9170 | 595 | 7387 | 668 | 0 |
| 2005 | 8771 | 0 | 3689 | 794 | 3353 | 1508 | 0 |  |  | 7485 | 1323 | 5588 | 2513 |  | 19154 | 0 | 3204 | 734 | 4815 | 351 | 0 | 20066 | 0 | 6515 | 1223 | 8025 | 585 | 0 |
| 2006 | 12762 | , | 3736 | 1800 | 2944 | 1455 |  | 3369 |  | 7584 | 2999 | 4907 | 2426 |  | 16704 | 0 | 3387 | 901 | 3945 | 403 | 0 | 17500 | 0 | 6904 | 1502 | 6575 | 672 | 0 |
| 2007 | 8515 | , | 3758 | 1710 | 1830 | 1024 |  |  |  | 7631 | 2850 | 3051 | 1707 | 0 | 14832 | 0 | 3638 | 1301 | 3171 | 305 | 0 | 15604 | 0 | 7406 | 2168 | 5285 | 508 | 0 |
| 2008 | 16445 | 0 | 5542 | 2266 | 3144 | 1401 |  |  | 0 | 1126 | 3776 | 5240 | 2336 |  | 15216 | 0 | 5187 | 1328 | 5423 | 390 | 0 | 16002 | 0 | 10595 | 2213 | 9038 | 649 | 0 |
| 2009 | 8872 | 0 | 3601 | 903 | 1907 | 1056 |  |  |  | 7306 | 1505 | 3178 | 1759 | 0 | 18479 | 0 | 3727 | 950 | 4556 | 275 | 0 | 19412 | 0 | 7589 | 1584 | 7594 | 458 | 0 |
| 2010 | 12889 | 0 | 4801 | 993 | 1675 |  |  |  |  | 9746 | 1655 | 2792 | 1802 |  | 21375 | 0 | 4488 | 1047 | 3656 | 338 | 0 | 22454 | 0 | 9157 | 1744 | 6093 | 564 | 0 |
| 2011 | 17993 | 0 | 5120 | 1365 | 4441 |  |  |  |  | 10386 | 2276 | 7402 | 2824 |  | 26977 | 0 | 4697 | 1571 | 6007 | 483 | 0 | 28373 | 0 | 9529 | 2619 | 10011 | 805 | 0 |
| 2012 | 9566 | 0 | 3615 | 584 | 3550 | 28 |  | 10038 |  | 7332 | 973 | 5916 | 2047 |  | 17918 | 0 | 3665 | 904 | 4488 | 313 | 0 | 18837 | 0 | 7434 | 1507 | 7481 | 522 | 0 |
| 2013 | 7219 | 88 | 3185 | 411 |  | 01 |  | 7574 | 104 | 6461 | 685 | 4111 | 2335 |  | 22026 | 205 | 4171 | 989 | 3938 | 339 | 0 | 23135 | 242 | 8461 | 1648 | 6563 | 565 | 0 |
| 2014 | 9193 | 0 | 3945 | 676 | 2412 |  |  | - 9647 | - 0 | 8003 | 1127 | 4020 | 3901 |  | 10954 | 0 | 2400 | 695 | 1593 | 1035 | 0 | 11504 | 0 | 4869 | 1159 | 2655 | 1725 | 0 |
| 2015 | 17189 | 0 | 5843 | 1253 | 3772 |  | 0 | 18048 | 0 | 11853 | 2088 | 6287 | 3046 |  | 18449 | 0 | 3995 | 1641 | 3943 | 580 | 0 | 19388 | 0 | 8104 | 2735 | 6572 | 967 | 0 |
| 2016 | 13183 | 0 | 5977 | 1049 | 3202 |  |  | 13850 | 0 | 12124 | 1749 | 5336 | 4593 |  | 17866 | 0 | 4244 | 1911 | 5022 | 1037 | 0 | 18785 | 0 | 8608 | 3186 | 8370 | 1728 | 0 |
| 2017 | 8111 | 0 | 4562 | 503 | 3673 |  |  | 8528 | 0 | 9254 | 839 | 6121 | 2978 |  | 19049 | 0 | 4639 | 1066 | 4926 | 522 | 0 | 20023 | 0 | 9411 | 1777 | 8211 | 870 | 0 |
| 2018 | 9367 | 0 | 4041 | 701 | 2825 | 1593 |  | 9847 | 0 | 8198 | 1169 | 4708 | 2655 |  | 13960 | 0 | 3544 | 995 | 2596 | 640 | 0 | 14675 | 0 | 7189 | 1658 | 4327 | 1066 |  |

Appendix 4.ix. (Continued). Input data for small salmon and large salmon spawners to Quebec by category of data used in the run-reconstruction.

| Year | $\underset{\llcorner }{\mathrm{QCSSmCl}}$ | QCSSmC2 _L | QCSSmC3 _L | QCsSmC4 _L | QCSSmC5 | QCSSmC6 _L | QCSSmC1 | $\begin{gathered} \text { QCSSmC2 } \\ \mathrm{H} \end{gathered}$ | $\begin{aligned} & \text { QCSSmC3 } \\ & \text { _H } \end{aligned}$ | QCSSmC4 _H | QCSSmC5 _H | $\begin{aligned} & \text { QCSSmC6 } \\ & \mathrm{H} \end{aligned}$ | $\begin{aligned} & \text { QCSLgC1 } \\ & \_\mathrm{L} \end{aligned}$ | $\begin{aligned} & \text { QCSLgC2 } \\ & \_L \end{aligned}$ | $\begin{aligned} & \text { QCSLgC3 } \\ & \text { L } \end{aligned}$ | $\begin{gathered} \text { QCSLgC4 } \\ \mathrm{L} \end{gathered}$ | QCSLgC5 | $\begin{aligned} & \text { QCSLgC6 } \\ & \text { _L } \end{aligned}$ | $\begin{aligned} & \text { QCSLgC1 } \\ & \text { H } \end{aligned}$ | $\begin{aligned} & \text { QCSLgC2 } \\ & \text { _H } \end{aligned}$ | $\begin{aligned} & \text { QCSLgC3 } \\ & \text { _H } \end{aligned}$ | $\begin{aligned} & \text { QCSLgC4 } \\ & \text { H } \end{aligned}$ | $\begin{aligned} & \text { QCSLgC5 } \\ & \text { _H } \end{aligned}$ | $\begin{aligned} & \text { QCSLgC6 } \\ & \text { H } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 3061 | 4342 | 1915 | 415 | 1264 | 5160 | 3316 | 4547 | 5013 | 747 | 2378 |  | 10421 |  | 1861 | 2357 | 2815 | 5071 | 11933 | 7935 | 4974 | 5427 | 5290 | 8452 |
| 1985 | 3960 | 1622 | 1025 | 209 | 4241 | 4384 | 4563 | 1687 | 2844 | 351 |  | 7307 |  |  | 2125 | 340 | 7214 | 3351 | 11581 | 5244 | 6098 | 572 | 13826 | 5586 |
| 1986 | 6337 | 3827 | 1499 | 63 | 5151 | 5133 | 7160 | 3955 | 3998 | 107 |  | 8555 | 13659 | 5804 | 3695 | 35 | 4498 | 4971 | 15672 | 6045 | 10139 | 61 | 8416 | 8284 |
| 1987 | 7493 | 3489 | 2365 | 291 | 6411 | 5501 | 8319 | 3624 | 6388 | 510 | 1200 | 9168 | 13432 | 4791 | 3025 | 246 | 4830 | 3012 | 14875 | 4965 | 8422 | 431 | 9053 | 5019 |
| 1988 | 8173 | 4188 | 2738 | 419 | 6432 | 6423 | 9227 | 4353 | 7296 | 731 | 12059 | 10 | 15535 | 4258 | 4381 | 312 | 5172 | 4781 | 17069 | 4444 | 12177 | 542 | 9698 | 7969 |
| 1989 | 7779 | 3810 | 1878 | 273 | 5149 | 5622 | 8568 | 3945 | 4894 | 475 | 9645 | 9369 | 14645 | 6742 | 3239 | 253 | 4375 | 4567 | 16273 | 6975 | 8961 | 440 | 8188 | 7611 |
| 1990 | 8735 | 5757 | 2822 | 604 | 5437 | 2976 | 9768 | 5936 | 7362 | 1068 | 10167 | 4960 | 12398 | 8463 | 4557 | 1228 | 3950 | 2424 | 13522 | 8701 | 13006 | 2139 | 7402 | 4040 |
| 1991 | 7247 | 4551 | 2465 | 316 | 3827 | 2001 | 7942 | 4687 |  | 551 | 7166 | 3336 | 14061 | 5019 | 3970 | 596 | 2940 | 357 | 15392 | 5223 | 11389 | 1061 | 5511 | 595 |
| 1992 | 5989 | 4841 | 2937 | 370 | 3957 | 3462 | 6648 | 5057 |  | 657 | 7378 | 5770 | 12850 | 4819 | 4492 | 325 | 2044 | 1503 | 14036 | 5019 | 12711 | 575 | 3835 | 2505 |
| 1993 | 4852 | 4311 | 2524 | 747 | 3339 | 1447 | 5592 | 4503 | 6626 |  |  | 2412 | 9848 | 6936 | 1809 | 282 | 2038 | 333 | 10724 | 8266 | 4792 | 533 | 3803 | 555 |
| 1994 | 5506 | 3996 | 2501 | 894 | 3089 | 437 | 6241 |  |  |  |  | 729 | 10468 | 5920 | 1693 | 448 | 2173 | 145 | 11524 | 6989 | 4460 | 787 | 4075 | 242 |
| 1995 | 5348 | 2835 | 1760 | 877 | 2956 | 434 | 5943 |  | 453 |  |  | 723 | 16562 | 8323 | 1321 | 781 | 3367 | 154 | 17872 | 9661 | 3310 | 1326 | 6294 | 256 |
| 1996 | 10636 | 1330 | 2260 | 372 | 3678 | 500 | 11748 |  | 6030 | 692 | 6828 | 833 | 16431 | 4417 | 2389 | 394 | 1924 | 135 | 17869 | 5536 | 6390 | 781 | 3593 | 225 |
| 1997 | 8238 | 142 | 2250 | 266 | 3074 | 462 |  |  |  | 461 | 5426 | 770 | 13433 | 3393 | 1744 | 308 | 2237 | 138 | 14481 | 4326 | 4696 | 528 | 4147 | 229 |
| 1998 | 7734 | 995 | 2347 | 289 | 4229 | 1124 |  |  | 16 | 516 | 7643 | 1875 | 10402 | 4429 | 1849 | 302 | 2213 | 290 | 11113 | 5464 | 4895 | 547 | 4073 | 484 |
| 1999 | 8155 | 509 | 2495 | 1653 | 4581 | 1426 |  |  | 5837 | 2883 | 8182 | 2379 | 14169 | 747 | 1962 | 3100 | 2956 | 491 | 14957 | 900 | 4640 | 5542 | 5536 | 837 |
| 2000 | 8291 | 372 | 693 | 519 | 5900 |  | 9040 |  |  | 921 | 12551 | 1005 | 11937 | 570 | 3322 | 491 | 5096 | 363 | 12669 | 620 | 7368 | 864 | 9376 | 749 |
| 2001 | 5329 | 143 | 1870 | 263 | 2579 |  | 5867 | 86 |  | 440 | 4729 | 1137 | 14527 | 505 | 4281 | 239 | 2980 | 348 | 15579 | 751 | 8986 | 400 | 5639 | 717 |
| 2002 | 9296 | 31 | 2231 | 658 | 3405 |  | 191 |  | 5572 | 1118 | 6294 | 2414 | 10843 | 8 | 4071 | 313 | 1500 | 344 | 11509 | 9 | 8833 | 539 | 2902 | 574 |
| 2003 | 8180 | 0 | 2269 | 661 | 2826 | 509 |  |  | 5604 | 1141 | 5204 | 2517 | 18832 | 0 | 5164 | 267 | 3763 | 383 | 19872 | 0 | 11156 | 447 | 7022 | 640 |
| 2004 | 9030 | 29 | 5574 | 278 |  | - 1639 |  |  | 12152 | 468 | 7222 | 2731 | 15558 | 107 | 4231 | 355 | 3268 | 401 | 16232 | 126 | 8531 | 593 | 6223 | 668 |
| 2005 | 6339 | 0 | 3025 | 716 |  |  |  |  | 6821 | 1245 | 4945 | 2511 | 16485 | 0 | 2901 | 719 | 3556 | 351 | 17397 | 0 | 6212 | 1208 | 6766 | 585 |
| 2006 | 8628 | 0 | 3159 | 1691 | 72 | 145 | 9235 |  | 7007 | 2890 | 4335 | 2426 | 14977 | 0 | 3055 | 872 | 2863 | 403 | 15773 | 0 | 6572 | 1473 | 5493 | 672 |
| 2007 | 5768 | 0 | 3226 | 1511 |  |  | 6217 | 0 | 7099 | 2651 | 2722 | 1707 | 12470 | 0 | 3203 | 1287 | 2444 | 303 | 13242 | 0 | 6971 | 2154 | 4558 | 506 |
| 2008 | 10562 | 0 | 4882 | 1756 |  | 1401 | 11467 | 0 | 10601 | 3266 | 4618 | 2336 | 13725 | 0 | 4676 | 1266 | 4296 | 390 | 14511 | 0 | 10084 | 2151 | 7911 | 649 |
| 2009 | 6293 | 0 | 3115 | 764 |  | 1056 | 6736 | 0 | 6820 | 1366 | 2904 | 1759 | 16489 | 0 | 3188 | 849 | 3588 | 275 | 17422 | 0 | 7050 | 1483 | 6626 | 458 |
| 2010 | 8860 | 0 | 4289 | 914 |  | 1080 | 9509 | 0 | 9234 | 1576 | 2428 | 1801 | 19170 | 0 | 3926 | 1023 | 3017 | 338 | 20249 | 0 | 8595 | 1720 | 5454 | 564 |
| 2011 | 12143 | 0 | 4496 | 1116 | 3674 | 1688 | 13049 | 0 | 9762 | 2027 | 6635 | 2818 | 24130 | 0 | 4180 | 1497 | 4579 | 479 | 25526 | 0 | 9012 | 2545 | 8583 | 801 |
| 2012 | 6620 | 0 | 3152 | 472 | 2924 | 1225 | 7092 | 0 | 6869 | 861 | 5290 | 2044 | 16098 | 0 | 3221 | 868 | 3685 | 313 | 17017 | 0 | 6990 | 1471 | 6678 | 522 |
| 2013 | 4959 | 88 | 2840 | 328 | 2131 | 1401 | 5314 | 104 | 6116 | 602 | 3776 | 2335 | 19804 | 205 | 3701 | 920 | 2925 | 338 | 20913 | 242 | 7991 | 1579 | 5550 | 564 |
| 2014 | 6579 | 0 | 3239 | 560 | 2127 | 2175 | 7033 | 0 | 7297 | 1011 | 3735 | 3735 | 10089 | 0 | - 2138 | 664 | 1206 | 740 | 10639 | 0 | 4607 | 1128 | 2268 | 1430 |
| 2015 | 11926 | 0 | 4709 | 986 | 2935 | 1820 | 12785 | 0 | 10719 | 1821 | 5450 | 3039 | 17003 | 0 | 3560 | 1579 | 3330 | 569 | 17942 | 0 | 7669 | 2673 | 5959 | 956 |
| 2016 | 9541 | 0 | 5006 | 887 | 2484 | 2591 | 10208 | 0 | 11153 | 1587 | 4618 | 4428 | 16613 | 0 | - 3875 | 1887 | 4103 | 1000 | 17532 | 0 | 8239 | 3162 | 7451 | 1691 |
| 2017 | 5810 | 0 | 3807 | 441 | 2806 | 1704 | 6227 | 0 | 8499 | 773 | 5254 | 2895 | 17991 | 0 | 0 4196 | 1042 | 3939 | 505 | 18965 | 0 | 8968 | 1753 | 7224 | 853 |
| 2018 | 6448 | 0 | 3373 | 581 | 2317 | 1543 | 6928 | 0 | 7530 | 1039 | 4200 | 2605 | 13394 | 0 | 3321 | 986 | 1996 | 626 | 14109 | 0 | 6966 | 1649 | 3727 | 1052 |

## Appendix 4.ix. (Continued). Input data for small salmon and large salmon used in the run-reconstruction



Appendix 4.x. Input data for 2SW, large, and small salmon returns to Salmon Fishing Areas 15 to 18 for Canada used in the run-reconstruction.

| Year | SF15R2_L | SF15R | SF16 | SF16 | SF17R2 | SF17R2 | SF18 | SF18R2 | SF15L | SF15Lg_H | SF16Lg_L | SF16Lg_H S | SF17Lg_L | SF17Lg_H S | SF18Lg_L | SF18Lg_H | SF15Sm | 55 m | F16 | 16sm | , | m | m | 8Sm_H |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 8243 | 10576 | 42901 | 45798 | 31 | 60 | 4744 | 6836 | 12681 | 16270 | 46462 | 49599 | 31 | 60 | 6161 | 7858 | 283 | 6279 | 47779 | 67697 | 0 | 0 | 264 | 1073 |
| 1971 | 3587 | 4616 | 26038 | 30669 | 29 | 29 | 1891 | 2782 | 5518 | 7102 | 28365 | 33409 | 29 | 29 | 2456 | 3198 |  | 4681 | 38388 | 54120 | 0 | 0 | 65 | 265 |
| 1972 | 4980 | 9756 | 29092 | 43510 | 402 | 402 | 4693 | 6024 | 8441 | 16536 | 30146 | 45087 | 402 | 402 | 6095 | 6924 | 2185 |  | 8886 | 69270 | 0 | 0 | 31 | 530 |
| 197 | 6211 | 1200 | 265 | 40492 | 206 | - 206 | 4140 | 5481 | 8393 | 16229 | 27771 | 42276 | - 206 | - 206 | 5376 |  | 3010 | 6668 | 190 | 6683 | 5 | 9 | 51 | 2095 |
| 1974 | 7264 | 14570 | 39270 | 60090 | 386 | 386 | 5481 | 6928 | 9950 | 19959 | 43249 | 66179 | 386 | 386 | 7119 | 7963 |  | 4895 | 8091 | 110470 | 0 | 0 | 187 | 757 |
| 1975 | 4353 | 7922 | 25889 | 39325 | 345 | 345 | 3452 | 4340 | 5510 | 10028 | 29826 | 45305 | 345 | 345 | 4483 |  |  | 5298 | 69993 | 98443 | 0 | 0 | 112 | 454 |
| 1976 | 7293 | 14416 | 20448 | 30758 | 575 | 578 | 2755 | 3674 | 9596 | 18969 | 23943 | 36016 | 575 | 578 |  |  |  | 析 | 96504 | 136107 | 14 | 28 | 299 | 1212 |
| 1977 | 9174 | 18077 | 49881 | 73330 | 606 | 606 | 85 | 5463 | 11053 | 21779 | 52673 | 77434 | 606 | 606 |  | 6280 | 6085 |  | 30621 | 42689 | 0 | 0 | 215 | 871 |
| 1978 | 545 | 1074 | 1950 | 2604 | 0 | 0 | 4585 | 6265 | 7277 | 14332 | 22653 | 30245 | 0 | 0 |  | 72 | 4350 | 7749 | 29783 | 39927 | 0 | 0 | 78 | 316 |
| 1979 | 1472 | 2535 | 6501 | 9306 | 459 | 463 | 1290 | 2014 | 2886 | 4971 | 9435 | 13507 | 459 | 463 |  |  | 4378 | 9495 | 50667 | 70714 | 2 | 5 | 1857 | 7536 |
| 1980 | 7102 | 14045 | 35163 | 48457 | 2 | - 5 | 3732 | 5177 | 8768 | 17340 | 37014 | 51008 | 2 | 2 | $5 \quad 48$ |  |  | 15278 | 41687 | 58839 | 12 | 23 | 520 | 2108 |
| 1981 | 4572 | 7357 | 11144 | 19268 | 40 | 77 | 2490 | 3769 | 9729 | 15652 | 16708 | 28887 | 40 |  | 3234 |  |  | 17119 | 63278 | 108226 | 259 | 498 | 2797 | 11348 |
| 1982 | 4314 | 6313 | 21442 | 41643 | 16 | 31 | 4135 | 5901 | 7311 | 10700 | 26504 | 51475 |  |  | 5370 |  | 6541 | 13383 | 78072 | 133171 | 175 | 336 | 2150 | 8722 |
| 1983 | 3453 | 5280 | 16349 | 28419 | 17 | 32 | 3733 | 5241 | 5852 | 8950 | 20309 | 35304 | -17 | 32 | 4848 |  | 2723 | 4638 | 24585 | 41332 | 17 | 32 | 21 | 858 |
| 1984 | 3329 | 6092 | 12216 | 31455 | 13 | 26 | 2391 | 3573 | 4214 | 7711 | 12941 | 33321 | 13 | 26 | 3105 | 410 | 12003 | 15867 | 28714 | 49595 | 17 | 32 | 460 | 1867 |
| 1985 | 4805 | 9500 | 14614 | 37625 | - 8 | 15 | 921 | 4481 | 7627 | 15080 | 16798 | 43247 |  | 15 | 1196 | 5150 | 7003 | 15516 | 53393 | 92224 | 113 | 217 | 730 | 3167 |
| 1986 | 831 | 15403 | 21617 | 55640 | 5 | 11 | 2274 | 11479 | 10305 | 20267 | 25342 | 65228 |  |  | 2953 | 13195 | 10813 | 23926 | 103230 | 178295 | 566 | 1088 | 965 | 3854 |
| 1987 | 836 | 23 | 1252 | 32224 | 66 | 128 | 2446 | 10156 | 7556 | 14255 | 15734 | 40483 | 66 |  |  | 11674 | 9630 | 21220 | 74485 | 128644 | 1141 | 2194 | 1541 | 8586 |
| 1988 | 7152 | 13998 | 14384 | 36938 | 96 | 185 | 2365 | 9851 | 9933 | 19441 | 17627 |  |  | -185 | 3071 | 11322 | 13168 | 29092 | 107071 | 184904 | 1542 | 2963 | 1297 | 7353 |
| 1989 | 4390 | 8492 | 9113 | 23385 | 149 | 287 | 1970 | 8288 | 7701 | 14898 | 13955 |  |  | 287 | 2558 | 9526 | 6357 | 13900 | 66069 | 114097 | 400 | 770 | 835 | 4843 |
| 1990 | 4326 | 8369 | 14269 | 36639 | 284 | 545 | 1778 | 7471 | 6362 | 12307 | 23164 | 594 |  |  | 2309 | 8588 | 7880 | 17314 | 73020 | 126115 | 1842 | 3539 | 921 | 5335 |
| 1991 | 2387 | 4668 | 1468 | 37736 | 188 | 361 | 2181 | 282 | 4773 | 9335 | 24273 | 6237 |  | 361 | 2832 | 10669 | 441 | 982 | 53453 | 92327 | 1576 | 3028 | 1089 | 6266 |
| 1992 | 4002 | 7787 | 2138 | 30728 | 95 | 183 | 2229 | 9314 | 7411 | 14420 | 34573 | 4968 |  | 183 | 2895 | 10706 | 8853 | 19614 | 142416 | 204708 | 1873 | 3599 | 936 | 5335 |
| 1993 | 1395 | 2684 | 15579 | 60246 | 22 | 43 | 1266 | 5193 | 3487 | 6711 | 22602 | 87407 | 22 | 43 | 1644 | 5969 | 5783 | 12812 | 70090 | 175096 | 1277 | 2454 | 1085 | 5900 |
| 1994 | 3960 | 7745 | 13652 | 24887 | 169 | 310 | 1866 | 7909 | 6600 | 12908 | 18098 | 32992 | 169 | 310 | 2423 | 9090 | 9136 | 20208 | 41773 | 59888 | 210 | 385 | 626 | 3640 |
| 1995 | 2713 | 5333 | 25593 | 37215 | 384 | 576 | 1395 | 5959 | 4171 |  | 30324 | 44094 | 384 | 576 | 1812 | 6850 | 2902 | 6429 | 44357 | 63453 | 658 | 987 | 509 | 3006 |
| 1996 | 3917 | 7754 | 11126 | 19117 | 394 | 591 | 2931 | 12652 | 6026 | 1929 | 3317 | 28035 | 394 | 591 | 3806 | 14542 | 6034 | 13370 | 32067 | 45995 | 710 | 1065 | 2241 | 13433 |
| 1997 | 2488 | 4898 | 8545 | 14244 | 387 | 581 | 3174 | 13848 |  |  |  |  | 387 | 581 | 4122 | 15917 | 5797 | 12845 | 14377 | 24122 | 517 | 776 | 518 | 3183 |
| 1998 | 1687 | 3260 | 6438 | 10199 | 385 | 577 | 1921 | 8397 |  | 5015 | 155 | 49 | 385 | 577 | 2494 | 9651 | 6288 | 13932 | 23035 | 32455 | 508 | 762 | 588 | 3631 |
| 1999 | 1780 | 3425 | 7175 | 10486 | 383 | 575 | 1457 | 6422 | 38 | 5 | 14744 | 21547 | 383 | 575 | 1892 | 7382 | 4936 | 10929 | 22006 | 28888 | 413 | 620 | 649 | 4011 |
| 0 | 2270 | 4410 | 7716 | 129 | 378 | 566 | 1336 |  |  |  | 16280 | 23823 | - 378 | 566 | 173 | 691 | 45 | 1652 | 3236 | 4125 | 395 | 593 | 55 | 3523 |
| 2001 | 3678 | 7239 | 14566 | 18907 | 376 | 564 | 1558 |  |  | 137 | 22560 | 29283 | 376 | 564 | 2023 | 7969 | 4792 | 10611 | 27497 | 36530 | 415 | 622 | 736 | 4610 |
| 2002 | 2234 | 4337 | 5671 | 8753 | 372 | 557 | 1134 | 138 |  | 73 | 11291 | 17429 | 372 | 557 | 1472 | 5905 | 11213 | 24838 | 41274 | 53641 | 390 | 585 | 757 | 4821 |
| 2003 | 3891 | 7667 | 11284 | 16297 | 371 | 557 |  | 9899 | - 5987 | 1795 | 19283 | 27850 | 371 | 557 | 2861 | 11378 | 3060 | 6773 | 27838 | 38840 | 515 | 773 | 703 | 4498 |
| 200 | 2951 | 5777 | 1105 | 17301 | 367 | 550 |  | 1019 | 40 |  | 19397 | 30353 | 367 | 50 | 3170 | 12665 | 11837 | 26220 | 44699 | 61570 | 330 | 495 | 988 | 633 |
| 2005 | 3359 | 6598 | 1185 | 19579 | 373 | 560 |  |  | 5168 | 10150 | 17827 | 29443 | 373 | 560 | 2649 | 10331 | 4043 | 8952 | 30330 | 47497 | 343 | 514 | 723 | 4472 |
| 2006 | 2516 | 4903 | 10299 | 16221 | 392 | 587 |  |  | 3870 | 7543 | 19807 | 31195 | 392 | 587 | 2539 | 10038 | 8618 | 19088 | 31729 | 49352 | 331 | 497 | 786 | 5003 |
| 2007 | 4258 | 8403 | 9592 | 14478 | 412 |  |  | 45 |  | 12928 | 16257 | 24539 | 412 | 618 | 1747 | 6949 | 4189 | 9275 | 23220 | 41115 | 275 | 413 | 650 | 4123 |
| 2008 | 2852 | 5580 | 5874 | 10639 | 429 |  |  |  | 4388 | 8584 | 10489 | 18999 | 429 | 644 | 2638 | 10683 | 13449 | 29791 | 26917 | 45734 | 298 | 447 | 1127 | 7400 |
| 2009 | 3964 | 7813 | 10919 | 16378 | 402 | 602 | 22 |  | 099 | 12020 | 17611 | 26416 | 402 | 602 | 1977 | 8222 | 5177 | 11464 | 11581 | 19568 | 233 | 350 | 267 | 1992 |
| 2010 | 2978 | 5831 | 8013 | 11308 |  | 658 | 2051 | 951 | 4581 | 8972 | 15711 | 22173 | 439 | 658 | 2664 | 10933 | 8020 | 17762 | 49067 | 65290 | 258 | 387 | 783 | 5330 |
| 2011 | 7265 | 14445 | 22570 | 39415 |  | 980 |  | 15787 | 11177 | 22223 | 26553 | 46371 | 653 | 980 | 4518 | 18146 | 9951 | 22040 | 42014 | 64198 | 291 | 436 | 1065 | 6965 |
| 2012 | 3230 | 6338 | 8561 | 14565 | 65 |  | 835 | - 4083 | 4969 | 9750 | 12230 | 20808 | 653 | 980 | 1084 | 4693 | 4251 | 9411 | 7693 | 14039 | 291 | 436 | 184 | 1501 |
| 2013 | 5324 | 10544 | 7643 | 14546 | - 719 |  | 1848 | 8498 | 8190 | 16222 | 11077 | 21081 | 719 | 1077 | 2400 | 9768 | 4559 | 10095 | 10448 | 20603 | 274 | 410 | 394 | 2743 |
| 2014 | 2704 | 5282 | 5903 | 11463 | 491 |  | 1250 | 5924 | 4160 | 8126 | 8758 | 17008 | 491 | 735 | 1623 | 6809 | 3697 | 8183 | 6787 | 11513 | 222 | 332 | 253 | 1915 |
| 2015 | 3410 | 6700 | 9559 | 17880 | 564 |  | 1752 | 8250 | 5247 | 10308 | 13959 | 26109 | 564 | 846 | 2276 | 9483 | 6135 | 13586 | 25404 | 34585 | 235 | 349 | 687 | 4731 |
| 2016 | 3052 | 5980 | 10833 | 22657 | 460 |  | 1640 | 7823 | 4695 | 9201 | 15051 | 31479 | 460 | 688 | 2130 | 8992 | 3921 | 8680 | 13094 | 22423 | 180 | 268 | 341 | 2544 |
| 2017 | 2416 | 4703 | 9242 | 16163 | 459 | 687 | 1183 | 5820 | 3717 | 7235 | 13322 | 23299 | 459 | 687 | 1537 | 6690 | 3570 | 7902 | 12127 | 19170 | 203 | 303 | 533 | 3722 |
| 2018 | 1923 | 3712 | 10651 | 21341 | 44 | 670 | 1447 | 7337 | 2958 | 571 | 15381 | 30819 | 7 | 670 | 1879 | 8434 | 4378 | 9692 | 7540 | 12810 | 185 | 277 | 357 | 2859 |

Appendix 4.x. (continued). Input data for 2SW, large, and small salmon spawners to Salmon Fishing Areas 15 to 18 for Canada used in the run-reconstruction.

| Year | SF15S2_L | SF15S2_H | SF16S2_L | SF16S2_ | SF17S2_L S | SF17S2_H | SF18S2_L | SF18S2 | SF15SLg | 15SLg_ | 6SLg_L | 6SLg | g_ | 7SL | BSLg_ | 8SLg | $15 S S m$ | 5SSm | 6SSm | 16SSm | Ssm | SSm | SSm | Sm |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 1156 | 3252 | 5346 | 8242 | 18 | 47 | 304 | 1587 | 1779 | 5003 | 5790 | 8926 | 18 | 47 | 395 | 1824 | 1417 | 4396 | 25958 | 45876 | 0 | 0 | 167 | 842 |
| 1971 | 510 | 1434 | 6724 | 11354 | 0 | 0 | 133 | 694 | 785 | 2207 | 7324 | 12369 | 0 | 0 | 173 | 797 |  | , | 22463 | 38195 | 0 | 0 | 41 | 208 |
| 1972 | 2367 | 6656 | 17031 | 31450 | 0 | 0 | 148 | 775 | 4011 | 11282 | 17648 | 32589 | 0 | 0 | 193 | 891 | 334 | 3208 | 7639 | 48023 | 0 | 0 | 82 | 416 |
| 1973 | 2873 | 8081 | 19277 | 33170 | 0 | 0 | 165 | 863 | 3883 | 10920 | 20126 | 34632 | 0 | 0 | 215 |  | 1505 | 4668 | 1703 | 51349 | 3 | 7 | 325 | 1645 |
| 1974 | 3620 | 10183 | 3119 | 52012 | 0 | 0 | 151 | 790 | 60 | 13949 | 34352 | 57282 | 0 | 0 | 196 | 908 |  | 3405 | 7376 | 89755 | 0 | 0 | 118 | 595 |
| 1975 | 1769 | 4975 | 18536 | 31972 | 0 | 0 | 91 | 473 | 2239 | 6297 | 21355 | 36834 | 0 | 0 | 118 |  |  | 3707 | 0438 | 78888 | 0 | 0 | 71 | 357 |
| 1976 | 3530 | 9928 | 11842 | 22152 | 1 | 4 | 116 | 604 | 4644 | 13063 | 13867 | 25940 | 1 | 4 | 151 | 694 | 2480 | 7692 | 64526 | 104130 | 8 | 22 | 188 | 951 |
| 1977 | 4412 | 12408 | 30623 | 54071 | 0 | 0 | 198 | 1033 | 5315 | 14949 | 32337 | 57097 | 0 | 0 |  | 187 |  |  | 13270 | 25338 | 0 | 0 | 135 | 684 |
| 1978 | 2622 | 7375 | 6998 | 3535 | 0 | 0 | 223 | 66 | 3496 | 833 | 8128 | 15720 | 0 | 0 |  |  | 1398 | 4337 | 14689 | 24833 | 0 | 0 | 49 | 248 |
| 1979 | 527 | 1482 | 3000 | 5806 | 3 | 7 | 115 | 598 | 1033 | 2906 | 4355 | 8426 | 3 | 7 |  |  | 2104 | 6528 | 31829 | 51876 | 1 | 4 | 1170 | 5915 |
| 1980 | 3440 | 9677 | 17667 | 30961 | 1 | 4 | 198 | 1033 | 4248 | 11947 | 18597 | 32590 | 1 | 4 |  |  |  | 9293 | 27791 | 44943 | 7 | 18 | 327 | 1655 |
| 1981 | 1380 | 3880 | 2392 | 10515 | 36 | 73 | 196 | 1027 | 2935 | 8256 | 3586 | 15765 | 36 |  | 255 |  |  | 9874 | 35423 | 80370 | 151 | 390 | 1762 | 8908 |
| 1982 | 991 | 2786 | 8418 | 28619 | 8 | 23 | 253 | 1322 | 1679 | 4723 | 10405 | 35376 | 8 |  | 329 |  | 038 | 9027 | 51324 | 106423 | 102 | 263 | 1354 | 6847 |
| 1983 | 906 | 2547 | 5516 | 17586 | 15 | 30 | 210 | 1100 | 1535 | 4317 | 6852 | 21846 | 15 | 30 | 273 |  | 820 | 2486 | 13298 | 30045 | 10 | 25 | 13 | 674 |
| 1984 | 2656 | 5402 | 11650 | 30889 | 13 | 26 | 259 | 1148 | 3362 | 6838 | 12341 | 32721 |  | 26 | 337 | 1320 | 1620 | 4971 | 7389 | 28271 | 10 | 25 | 177 | 1200 |
| 1985 | 4514 | 9180 | 14019 | 37030 | - 8 | 15 | 871 | 4359 | 7164 | 14571 | 16114 | 42563 |  | 15 | 1131 | 5010 | 3557 | 10936 | 32275 | 71106 | 66 | 170 | 145 | 1788 |
| 1986 | 7279 | 14804 | 20606 | 54630 | 5 | 11 | 2164 | 11213 | 9577 | 19479 | 24157 | 64044 |  |  | 2811 | 12889 | 5589 | 16990 | 71918 | 146983 | 330 | 852 | 63 | 1729 |
| 1987 | 4122 | 8383 | 11414 | 31114 | 66 | 128 | 2370 | 9923 | 6441 | 13099 | 14340 | 39088 | 66 |  | 307 | 11406 | 4867 | 14920 | 49971 | 104131 | 665 | 1718 | 422 | 4394 |
| 1988 | 658 | 13386 | 1380 | 3635 | 96 | 185 | 2283 | 9600 | 9141 | 18592 | 16913 |  | 96 |  | 2965 | 11035 | 6664 | 20468 | 71967 | 149800 | 899 | 2320 | 26 | 3467 |
| 1989 | 3944 | 8021 | 8466 | 22739 | 149 | 287 | 1903 | 8085 | 6919 | 14072 | 12965 |  |  | 287 | 2471 | 9293 | 3191 | 9741 | 37696 | 85724 | 233 | 603 | 174 | 2368 |
| 1990 | 3886 | 7903 | 13669 | 36039 | 284 | 545 | 1715 | 7279 | 5715 | 11623 | 22190 | 585 |  |  | 2227 | 8367 | 3996 | 12190 | 46902 | 99996 | 1074 | 2771 | 167 | 2510 |
| 1991 | 2193 | 4460 | 14200 | 37251 | 188 | 361 | 2106 | 9053 | 4386 | 8920 | 23472 | 6157 |  | 361 | 2735 | 10406 | 2215 | 6872 | 39648 | 78522 | 919 | 2371 | 199 | 2933 |
| 1992 | 3639 | 7400 | 20770 | 30116 | 95 | 183 | 2146 | 9062 | 6738 | 13704 | 33583 | 48697 |  | 183 | 2787 | 10416 | 4426 | 13728 | 116657 | 178949 | 1092 | 2818 | 13 | 2320 |
| 1993 | 1239 | 2521 | 15239 | 59907 | 22 | 43 | 1220 | 5055 | 3099 | 6302 | 22109 | 86914 | 22 | 43 | 1585 | 5811 | 2891 | 8968 | 52050 | 157056 | 745 | 1922 | 200 | 2583 |
| 1994 | 3639 | 7401 | 13418 | 24653 | 166 | 307 | 1805 | 7722 | 6065 | 123 | 17787 | 32682 | 166 | 307 | 2344 | 8876 | 4554 | 14125 | 25649 | 43764 | 118 | 292 | 135 | 1798 |
| 1995 | 2519 | 5124 | 25326 | 36949 | 380 | 576 | 1350 | 5821 | 3873 |  | 30007 | 43778 | 380 | 576 | 1753 | 6691 | 1451 | 4501 | 34650 | 53746 | 250 | 375 | 114 | 1527 |
| 1996 | 3688 | 7502 | 10743 | 18662 | 88 | 591 | 2850 | 12407 | 5674 | 11541 | 755 | 27367 | 388 | 591 | 3701 | 14261 | 3017 | 9359 | 19511 | 29260 | 258 | 387 | 81 | 8090 |
| 1997 | 2316 | 4710 | 8106 | 13754 | 385 | 581 | 3086 | 13582 |  |  |  |  | 385 | 581 | 4008 | 15611 | 2899 | 8991 | 8702 | 15524 | 256 | 384 | 160 | 1841 |
| 1998 | 1512 | 3076 | 6242 | 9970 | 382 | 577 | 1865 | 8227 |  | 4732 | 1508 | 95 | 382 | 577 | 2422 | 9457 | 3144 | 9752 | 14745 | 21339 | 255 | 382 | 183 | 2113 |
| 1999 | 1581 | 3217 | 6666 | 9947 | 379 | 575 | 1423 | 6320 | 433 | 4948 | 13697 | 20439 | 379 | 575 | 1848 | 7264 | 2465 | 7646 | 12552 | 17370 | 253 | 380 | 291 | 2669 |
| 2000 | 2057 | 4184 | 7417 | 10960 | 376 | 566 | 1307 |  |  |  | 15648 | 23122 | 376 | 566 | 1698 | 6816 | 3727 | 11560 | 19148 | 25367 | 252 | 378 | 254 | 2393 |
| 2001 | 3423 | 6962 | 13994 | 18296 | 374 | 564 | 1522 | 25 |  | 10711 | 21674 | 28337 | 374 | 564 | 1977 | 7845 | 2393 | 7423 | 16404 | 22727 | 250 | 376 | 31 | 3040 |
| 2002 | 2022 | 4112 | 5430 | 8485 | 371 | 557 | 1108 | 060 |  | 6327 | 10813 | 16895 | 371 | 557 | 1439 | 5817 | 5604 | 17382 | 26048 | 34704 | 249 | 373 | 340 | 3256 |
| 2003 | 3630 | 7382 | 10883 | 15851 | 368 | 557 | 2156 | 9755 | 558 | 357 | 18598 | 27087 | 368 | 557 | 2800 | 11212 | 1527 | 4737 | 16643 | 24344 | 248 | 371 | 327 | 3088 |
| 2004 | 2717 | 5527 | 10600 | 16789 | 365 | 550 |  | 10841 | 480 |  | 18597 | 29454 | 365 | 550 | 3094 | 12460 | 5916 | 18350 | 28445 | 40255 | 246 | 369 | 391 | 4106 |
| 2005 | 3113 | 332 | 11294 | 18949 | 371 | 560 |  |  | 4790 | 9742 | 16984 | 28495 | 371 | 560 | 2576 | 10137 | 2019 | 6262 | 18387 | 30404 | 246 | 368 | 242 | 2669 |
| 2006 | 2295 | 4668 | 9852 | 15720 | 390 | 587 | 1903 |  |  | 7181 | 18946 | 30231 | 390 | 587 | 2471 | 9855 | 4306 | 13357 | 20504 | 32840 | 247 | 370 | 276 | 3092 |
| 2007 | 3985 | 8105 | 9103 | 13945 | 409 |  | 31 |  |  | 12469 | 15428 | 23636 | 409 | 618 | 1705 | 6834 | 2092 | 6488 | 14547 | 27074 | 248 | 372 | 260 | 2661 |
| 2008 | 2622 | 5332 | 5439 | 10161 | 429 |  |  |  |  | 8204 | 9712 | 18145 | 429 | 644 | 2557 | 10465 | 6722 | 20850 | 17135 | 30307 | 249 | 373 | 273 | 4200 |
| 2009 | 3700 | 7526 | 10397 | 15808 | 401 | 602 |  |  | 5693 | 11578 | 16770 | 25496 | 401 | 602 | 1922 | 8074 | 2586 | 8020 | 6400 | 11991 | 233 | 350 | 69 | 1251 |
| 2010 | 2744 | 5580 | 7542 | 10808 |  | 658 | 1990 | 9324 | 4221 | 8585 | 14788 | 21192 | 438 | 658 | 2584 | 10718 | 4007 | 12429 | 31167 | 42523 | 256 | 384 | 192 | 3118 |
| 2011 | 6906 | 14043 | 21787 | 38480 |  | 980 | 3387 | 15507 | 10624 | 21604 | 25631 | 45271 | 652 | 980 | 4398 | 17825 | 4972 | 15424 | 27016 | 42545 | 290 | 435 | 365 | 4343 |
| 2012 | 2990 | 6079 | 8171 | 14121 |  |  | 819 | - 4036 | 4599 | 9352 | 11673 | 20173 | 652 | 980 | 1064 | 4639 | 2122 | 6584 | 4003 | 8445 | 290 | 435 | 92 | 1156 |
| 2013 | 5022 | 10211 | 7057 | 13898 | 715 |  | 1803 | 8362 | 7726 | 15709 | 10228 | 20141 | 715 | 1074 | 2342 | 9612 | 2277 | 7063 | 5599 | 12708 | 272 | 408 | 152 | 1837 |
| 2014 | 2478 | 5040 | 5813 | 11323 | 487 |  | 1228 | 5858 | 3812 | 7754 | 8624 | 16800 | 487 | 732 | 1595 | 6734 | 1845 | 5724 | 4124 | 7432 | 220 | 330 | 139 | 1488 |
| 2015 | 3163 | 6433 | 9387 | 17632 | 561 |  | 1727 | 8173 | 4866 | 9897 | 13707 | 25748 | 561 | 842 | 2242 | 9394 | 5945 | 13336 | 24170 | 33268 | 229 | 343 | 420 | 3730 |
| 2016 | 2815 | 5726 | 10571 | 22289 | 456 |  | 1615 | 7747 | 4331 | 8809 | 14687 | 30967 | 456 | 685 | 2097 | 8904 | 3797 | 8518 | 12213 | 21458 | 176 | 264 | 213 | 2066 |
| 2017 | 2198 | 4471 | 9033 | 15892 | 455 | 684 | 1162 | 5755 | 3382 | 6879 | 13021 | 22909 | 455 | 684 | 1509 | 6615 | 3457 | 7754 | 11289 | 18268 | 201 | 301 | 281 | 2777 |
| 2018 | 1720 | 3498 | 10479 | 21073 | 444 | 667 | 1423 | 7266 | 2646 | 5382 | 15133 | 30433 | 444 | 667 | 1848 | 8351 | 4241 | 9512 | 7273 | 12496 | 183 | 275 | 188 | 2225 |

Appendix 4.xi. Input data for 2SW, large, small salmon returns and spawners to Salmon Fishing Areas 19 to 21 for Canada used in the run-reconstruction.

| Year | SF19_21R:S | SF19_21RiS | F23R2_L S | SF23R2_H S | SF19_21L̨S | SF19_21L¢ | SF23Lg_L | SF23Lg_H | SF19_21Sr | SF19_21Sr | SF23Sm_LS | SF23Sm_HS | SF19_21S2S | SF19_21S2 | SF23S2_L | SF23S2_H | SF19_21SLS | SF19_21SL | SF23SLg_L | F23SLg_-S | SF19_21SSS | F19_21Ss | 23S5m_S | SF23SSm_ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 5600 | 7447 | 8540 | 12674 | 7273 | 9671 | 9691 | 13945 | 16177 | 24106 | 5306 | 7521 | 2388 | 4234 | 1536 | 4846 |  | 5499 | 1451 | 5705 | 9429 | 17358 | 3886 | 6101 |
| 1971 | 4120 | 5215 | 7155 | 10536 | 5350 | 6773 | 8056 | 11573 | 11911 | 18004 | 3248 | 4541 | 1418 | 2513 | 3612 | 6576 |  | 328 | 3888 | 7405 | 7246 | 13339 | 1216 | 2509 |
| 1972 | 5744 | 6993 | 7869 | 11368 | 7460 | 9082 | 8890 | 12536 | 11587 | 17992 | 1831 | 2506 | 1616 | 2865 | 6472 | 9806 | 099 |  | 7246 | 10892 | 7616 | 14021 | 0 | 1 |
| 1973 | 6922 | 8659 | 4205 | 6036 | 8049 | 10069 | 4760 | 6638 | 14169 | 22159 | 5474 | 7012 | 2246 | 3984 | 2752 | 4412 |  | 4632 | 050 | 4928 | 9502 | 17492 | 4037 | 5575 |
| 1974 | 13138 | 15363 | 10755 | 14988 | 13138 | 15363 | 12187 | 16444 | 25032 | 39058 | 10195 | 12901 | 2878 | 5103 | 8123 | 12046 |  | 5103 | 090 | 13347 | 16680 | 30706 | 8071 | 10777 |
| 1975 | 12261 | 13797 | 13107 | 18578 | 12261 | 13797 | 14829 | 20351 | 10860 | 15753 | 18022 | 23101 | 1987 | 3523 | 10987 |  |  | 3523 | 2335 | 17857 | 5819 | 10712 | 15363 | 20442 |
| 1976 | 8607 | 10104 | 1427 | 20281 | 8873 | 10416 | 16128 | 22175 | 21071 | 33009 | 22835 | 28864 | 1935 | 3432 | 10071 | 5583 |  |  | 11183 | 17230 | 14196 | 26134 | 17572 | 23601 |
| 1977 | 10872 | 12851 | 1686 | 2399 | 14119 | 1669 | 19165 | 2618 | 24599 | 37314 | 13738 | 16671 | 2559 | 4539 | 120 | 18568 |  |  | 13452 | 20470 | 15120 | 27835 | 9196 | 12129 |
| 1978 | 8272 | 9779 | 8225 | 11294 | 10471 | 12378 | 9335 | 12342 | 7621 | 10023 | 6271 | 7695 | 1948 | 3455 |  | 8076 | 2466 | 4373 | 5948 | 8955 | 2857 | 5259 | 4256 | 5680 |
| 1979 | 3781 | 4879 | 5165 | 7207 | 5180 | 6684 | 5856 | 7903 | 24298 | 37514 | 15356 | 20517 | 1419 | 2517 |  |  | 194 | 3448 | 4217 | 6264 | 15716 | 28932 | 11640 | 16801 |
| 1980 | 14094 | 17318 | 19056 | 26865 | 16388 | 20137 | 21464 | 29480 | 34377 | 50250 | 25139 | 31483 | 4170 | 7394 |  |  |  | 8598 | 13190 | 21206 | 18876 | 34749 | 19597 | 25941 |
| 1981 | 8662 | 11471 | 11026 | 15267 | 11706 | 15501 | 12481 | 16743 | 31204 | 48945 | 16826 | 21803 | 3631 |  | 3642 | 14 |  | 8702 | 3794 | 8056 | 21096 | 38837 | 7805 | 12782 |
| 1982 | 4458 | 5353 | 9782 | 13871 | 9485 | 11390 | 11147 | 15303 | 17619 | 27075 | 11811 | 15636 | 1158 | 2053 | 4475 |  | 464 | 4369 | 4903 | 9059 | 11244 | 20700 | 6532 | 10357 |
| 1983 | 4134 | 5356 | 9662 | 13836 | 6562 | 8501 | 10908 | 15235 | 9313 | 14068 | 9270 | 12592 | 1579 | 2800 | 468 |  | 2506 | 4445 | 92 | 4419 | 5653 | 10408 | 5132 | 8454 |
| 1984 | 1758 | 2854 | 15706 | 22627 | 2408 | 3909 | 17706 | 24992 | 18382 | 29867 | 15556 | 21678 | 416 | 2512 | 12280 | 8798 | 1940 | 3441 | 13675 | 20961 | 13658 | 25143 | 10290 | 16412 |
| 85 | 6894 | 12124 | 16541 | 23828 | 12 | 14968 | 8582 | 26289 | 484 | 954 | 305 | 17928 |  | 199 | 11885 | 824 | 4 | 14803 | 1310 | 20811 | 8024 | 318 | 8164 | 3036 |
| 1986 | 6755 | 11878 | 9891 | 14261 | 10722 | 18854 | 11142 | 15761 | 24369 | 39663 | 14274 | 20183 | 6624 | 748 | 7224 | 11280 | 10515 | 18647 | 8004 | 12623 | 18187 | 33481 | 10725 | 16634 |
| 1987 | 3748 | 6591 | 6922 | 10043 | 5950 | 10462 | 7865 | 11116 | 27269 | 44266 | 13358 | 17662 | 3676 |  |  | 8597 | 5835 | 10347 | 6343 | 9594 | 20213 | 37210 | 10257 | 14561 |
| 1988 | 4393 | 7735 | 4716 | 6697 | 7321 | 12891 | 5360 | 7312 | 24509 | 39750 | 16381 |  | 4322 | 7664 | 3420 | 5248 | 7203 | 12773 | 3835 | 5787 | 18125 | 33366 | 13061 | 19764 |
| 1989 | 4808 | 8469 | 6560 | 9437 | 696 | 12275 | 7393 | 10380 | 25602 | 41557 | 17579 | 2452 |  | 8396 | 6310 | 9158 | 6862 | 12168 | 7099 | 10086 | 18973 | 34928 | 13124 | 20066 |
| 1990 | 359 | 632 | 548 | 7918 | 619 | 1089 | 235 | 8710 | 2947 | 48039 | 13820 |  |  | 6260 | 4926 | 729 | 608 | 1079 | 5576 | 805 | 2208 | 4064 | 10025 | 15381 |
| 1991 | 2960 | 5213 | 7337 | 10563 | 4112 | 7240 | 8312 | 11659 | 9762 | 15955 | 13041 | 1768 |  | 5165 | 6080 | 9158 | 4045 | 7173 | 6833 | 10180 | 7363 | 13556 | 9495 | 14139 |
| 1992 | 2633 | 4634 | 6878 | 9809 | 3657 | 6437 | 7749 | 10726 | 13754 | 22269 | 13563 | 1840 | 588 | 4589 | 5826 | 8633 | 3594 | 6374 | 6511 | 9488 | 10125 | 18640 | 9485 | 14326 |
| 1993 | 2542 | 4470 | 4345 | 4820 | 3218 | 5658 | 5260 | 5980 | 13297 | 21681 | 7610 | 8828 | 2493 | 4421 | 3291 | 3654 | 3156 | 5596 | 4026 | 4746 | 9970 | 18354 | 5762 | 6868 |
| 1994 | 1360 | 2396 | 3084 | 3495 | 1743 | 3071 | 3659 | 4155 | 3154 |  | 5770 | 6610 | 1339 | 2375 | 2387 | 2680 | - 1717 | 45 | 2827 | 3273 | 2661 | 4900 | 4965 | 5738 |
| 1995 | 2253 | 3969 | 3439 | 3998 | 2532 | 446 | 3728 | 4289 | 8397 | 1387 | 265 | 9458 | 2218 | 3934 | 3126 | 3652 | 2492 | 4420 | 3362 | 3923 | 6512 | 11988 | 8025 | 9218 |
| 1996 | 3000 | 5278 | 4729 | 5397 | 3571 | 6283 | 5535 | 6365 | 1312 | 2293 | - 12907 | 15256 | 2946 | 5224 | 4009 | 4585 | - 3507 | 6219 | 4688 | 5497 | 10909 | 20082 | 11576 | 13892 |
| 1997 | 1163 | 2045 | 2769 | 3176 | 1550 | 2726 | 3210 | 3678 |  |  |  |  | 1140 | 2022 | 2219 | 2565 | 51520 | 2696 | 2565 | 3028 | 2917 | 5370 | 3971 | 4433 |
| 1998 | 924 | 1270 | 1372 | 1642 | 1359 | 1867 | 2032 | 2437 |  | 11927 | , | 801 | 915 | 1261 | 1068 | 1302 | 21346 | 1854 | 1675 | 2074 | 8818 | 11912 | 8775 | 10348 |
| 1999 | 1419 | 1951 | 2375 | 264 | 1709 | 23 | 34 |  | 71 | 5337 | 5508 | 6366 | 1409 | 1941 | 1934 | 2181 | 1697 | 2338 | 2251 | 2601 | 3895 | 5261 | 5196 | 6048 |
| 2000 | 1078 | 1483 | 98 | 1206 | 1315 | 1809 | 1189 |  |  |  | 4796 | 5453 | 1072 | 1477 | 805 | 1004 | 41307 | 1801 | 975 | 1216 | 6148 | 8305 | 445 | 5087 |
| 2001 | 1822 | 2506 | 1938 | 2279 | 1980 | 2724 | 2113 |  |  |  | 2513 | 2862 | 1812 | 2497 | 1699 | 2008 | 81970 | 2714 | 1831 | 2210 | 2315 | 3127 | 2210 | 2530 |
| 2002 | 382 | 525 | 483 | 548 | 749 | 1029 |  | 752 |  | 015 | 3501 | 3991 | 378 | 521 | 317 | 356 | 741 | 1021 | 442 | 542 | 5180 | 6998 | 3232 | 3689 |
| 2003 | 1854 | 2548 | 1056 | 1198 | 1952 | 2682 | 1128 | 1289 | - 2844 | 837 | 2292 | 2716 | 1834 | 2528 | 878 | 998 | 1931 | 2661 | 919 | 1074 | 2829 | 3822 | 2069 | 2469 |
| 2004 | 1028 | 1413 | 1335 | 1605 | 1302 | 1789 | 1402 |  | 847 |  | 3454 | 4297 | 1017 | 1401 | 1238 | 1492 | 21287 | 1774 | 1287 | 1574 | 3833 | 5178 | 3229 | 4039 |
| 2005 | 662 | 906 | 809 | 1012 | 860 | 1177 |  |  | 2870 | 3871 | 3597 | 4640 | 646 | 890 | 726 | 914 | -839 | 1156 | 791 | 1012 | 2854 | 3855 | 3433 | 4450 |
| 2006 | 1263 | 1734 | 922 | 1171 | 1559 | 2141 |  |  |  | 6940 | 3720 | 4743 | 1248 | 1720 | 796 | 1023 | 31541 | 2123 | 847 | 1113 | 5119 | 6915 | 3528 | 4501 |
| 2007 | 603 | 825 | 616 | 736 | 701 |  | - 689 | 41 |  | 5664 | 2466 | 3136 | 587 | 809 | 530 | 633 | 683 | 941 | 586 | 726 | 4176 | 5642 | 2305 | 2937 |
| 2008 | 1793 | 2465 | 812 | 1042 | 1928 |  |  |  | 82 | 9831 | 5924 | 7691 | 1778 | 2450 | 736 | 953 | 1912 | 2634 | 767 | 1007 | 7252 | 9801 | 5729 | 7467 |
| 2009 | 827 | 1135 | 1485 | 1886 | 1034 | 1418 | 678 |  | 2066 | 2788 | 1603 | 2027 | 811 | 1118 | 1391 | 1774 | 41014 | 1398 | 1565 | 2034 | 2051 | 2773 | 1472 | 1864 |
| 2010 | 934 | 1277 | 829 | 992 |  | 1451 | 1117 |  | 3686 | 4975 | 9114 | 11994 | 910 | 1253 | 726 | 877 | - 1034 | 1424 | 996 | 1275 | 3674 | 4963 | 9032 | 11901 |
| 2011 | 1489 | 2044 | 2486 | 3259 |  |  | 598 | 3421 | 3615 | 4878 | 4466 | 5943 | 1467 | 2023 | 2430 | 3196 | - 1482 | 2043 | 2532 | 3353 | 3601 | 4864 | 4391 | 5867 |
| 2012 | 661 | 903 | 268 | 331 |  |  | 335 | 422 | 350 | 470 | 178 | 219 | 641 | 883 | 238 | 298 | 761 | 1048 | 300 | 387 | 343 | 463 | 167 | 208 |
| 2013 | 2075 | 2852 | 420 | 543 | 2184 | 3002 | 503 | 660 | 922 | 1244 | 894 | 1151 | 2057 | 2834 | 405 | 526 | - 2165 | 2983 | 486 | 643 | 919 | 1241 | 870 | 1127 |
| 2014 | 415 | 569 | 172 | 217 | 415 |  | 230 | 299 | 529 | 713 | 677 | 910 | 410 | 564 | 163 | 208 | 8410 | 564 | 220 | 289 | 527 | 711 | 669 | 900 |
| 2015 | 338 | 463 | 248 | 314 | 377 |  |  | 326 | 1963 | 2650 | 1677 | 2131 | 332 | 457 | 244 | 309 | 9371 | 511 | 252 | 321 | 1960 | 2647 | 1654 | 2108 |
| 2016 | 883 | 1214 | 415 | 515 | 883 | 1214 | 457 | 568 | 284 | 383 | 1922 | 2528 | 878 | 1209 | 399 | 496 | $6 \quad 878$ | 1209 | 438 | 548 | 282 | 381 | 189 | 2494 |
| 2017 | 494 | 679 | 487 | 620 | 514 | 706 | 517 | 660 | 2357 | 3182 | 1003 | 1293 | 488 | 672 | 472 | 601 | - 507 | 699 | 499 | 638 | 2351 | 3176 | 995 | 1285 |
| 2018 | 1058 | 1456 | 162 | 203 | 1139 | 1567 | 168 | 213 | 455 | 614 | 748 | 875 | 1051 | 1448 | 160 | 201 | 1131 | 1559 | 166 | 211 | 453 | 612 | 743 | 870 |

## Appendix 4.xii. Input data for 2SW, large and small salmon returns and spawners to USA used in the run-reconstruction.



## Appendix 5: Model Walkthroughs

Summaries of the data preparation, model running and output processing were presented at a one-day workshop prior to the 2014 WGNAS meeting and provided step by step walkthroughs of the assessment processes. Where appropriate these have been updated in 2019.

## NEAC pre-fishery abundance and national conservation limit model in $R$

[NB: Instructions apply to model version: "NEAC_PFA_CL_RR_model_2015varM_v12" as used in 2015]

1) Introduction

This program performs the run-reconstruction estimation of pre-fishery abundance (PFA) of maturing and non-maturing 1SW salmon for each country (and region) in the NASCO-NEAC area. PFA is estimated for January 1st in the first sea winter. The program also establishes the pseudo stock-recruitment ( $S-R$ ) relationship between lagged egg deposition and Total 1SW PFA, and applies a hockey-stick S-R analysis to estimate the National/Regional Conservation Limit where river-specific CLs are not available. The original model is described by Potter et al. (2004); minor changes to the estimation approach used for different countries and regions have been reported in the annual reports of WGNAS.

The model also estimates the proportion (mean and SE) of the Faroes catch originating in different countries/regions based on the genetic analysis of scales collected in the fishery between 1993 and 1995 and the estimated PFA for each country/region since 2001, when no fishery has been operating at Faroes. This requires the model to be run once to provide the eatch proportions and then run a second time to provide a full PFA assessment. The catch proportions, are also used in the Catch Options model.

2 )
2.1 ) Load RStudio or R;
2.2) Set upa folder from which you will run the program;
2.3. Use folder and file names without spaces;
2.4 ) Put the program, the input files (annual and multiannual) and the summary data file (see 6f) in this folder.

3 ) Input Data
3.1 ) Annual data (filenames: Annual-data-XX-YY.txt)
3.1.1 ) There is a file for each country ( $X X$ ) and region ( $Y Y$ ) which contains the $40+$ year time-series of data on catches, exploitation rates and non-reporting rates (plus additional data for some countries).
3.1.2 ) To read the .txt files, it is easiest to open them from within Excel. i.e.

- Open Excel;
- select the correct folder;
- click on 'Open'
- You will probably need to change the setting in the lower right corner of the open box from 'Excel files' to 'All files';
- Double-click on the file you want to open and it should open the 'Text Import Wizard';
- select 'finish' (If this doesn't work reopen the file, but select 'Delimited' at step 1, 'Tab' at step 2 and 'General' at step 3.)
3.1.3 ) Do not add any formatting to the file. If loading a new version of a file that has been saved in Excel (e.g. after addition of a new year's data), re-save the file by clicking 'Save As' and selecting 'Text (Tab delimited)' from the 'Save as type' list. This will remove the formatting and add the
3.1.4 ) Do not change the file name.
3.1.5 ) Close and save the file before running the programme. You will be prompted to confirm that you want to lose the formatting; click 'yes'.
3.2 ) Multiannual-data (file-name: 'Multiannual-data.txt' or similar)
3.2.1) This file contains most of the other parameters used in the model including: smolt age composition, fecundity and sex ratios by region, M , etc.
3.2.2 ) The second value listed is the 'lastdatayear' which needs to be updated to the latest year for which data are provided in the Annual-data-XX files.
3.2.3 ) The file is not formatted in columns so can be read easily in Notebook, which should be selected automatically if you click on the file to open it. (NB: If you open the file in Excel, don't save it because it will probably add " " marks to each line.
3.2.4) All blank lines and lines starting with '\#' are ignored in this file. Apart from these:
- The first line must start with 'list(

The last line must be ')'
A1 other lines must be 'variable name' <- number, followed by a comer (except for the last data line which has no comer).
If the module estimating the composition of the Faroes catch is run (see below) the new values must be inserted at the end of the multiannual data file in place of the current ones.
3.2.6 ) Save the file before running the model.

4 ) Model structure
4.1 ) Introductory section: contains working directory, source files and various parameters controlling the way the program runs (some of these will need to be changed for your laptop (see below).
4.2 ) Functions: functions are sections of code that the program calls up to repeat the same job. They have to be run before they are first called by the program; this is achieved by placing them at the beginning of the code. The main functions run the hockey-stick analysis for the NCL model and output certain figures and tables.
4.3 ) Faroes and Greenland sections: these sections calculate the harvest in the distant water fisheries
4.4 ) NEAC country/regions sections: there is a section for each country (in alphabetic order) and region to calculate the main outputs of the R-R model.
4.5 ) Output summaries: this section creates NEAC summary figures and tables and the country/region data files for the Winbugs Forecast Model.
4.6 ) Faroes catch composition: The final section estimates the proportion (mean and SE) of the Faroes catch originating in different countries/regions. This requires the model to be run once to provide the catch proportions and then run a second time to provide a full PFA assessment.

5 ) Running the code from RStudio
5.1) Open R Studio
5.2 ) Select "File/Open File" and use the browser to select and open the code file; the code should open in the Top left panel. The code will have a name like "NEAC_PFA_CL_RR_model_2015_xxy
5.3 ) If you have been using the code recently, you can select "File/Recent Files" and select the file from the drop-down list (iffit is there); you can open several code files simultaneously and they appear as tabs above the Top Left panel.
5.4 ) To set up the code for your PCAlaptop, R-Click on the code and scroll down to:
line 40 -enter the full path name of the working directory (replace the text between the parentheses with the full pathname of the folder containing the code an your laptop (e.g. "D:/Modelling_NEAC/PFA_NCL_R/2014"). line 45 -ensure that the text between the parentheses shows the correct filename for the multi-annual data file.
s $77-86$-select which countries you wish to run the assessment for by setting "run-XX": $1=$ run country $X X ; 0=$ do not run. The summaries will only be run if all countries are set to 1 .
line 82 -set "PrintFigs" equal to ' 1 ' to output the summary figures (or any other value not to output them); otherwise " 0 ".
line 89 -set "WinbugsFiles" equal to ' 1 ' to output the data files for the Bayesian forecast model (or any other value not to output them) ; otherwise " 0 "..
line 92 -set "PrintCountryTables" equal to ' 1 ' to output summary output data for each region that is run (or any other value not to output them) ; otherwise " 0 ".
line 98 -set "RunFaroeseCatchSplitEstimation" equal to "TRUE" to run the estimation of the Faroes catch composition; otherwise "FALSE".
5.5 ) You do not need to save your changes before you run the code. [If you wish to save any changes, use "File/Save" or "File/Save As" as
normal. It's a good idea to include the extension ".R". NB: You will be prompted to save the file before you close it.]
5.5.1 ) To clear the 'console' area (lower left panel) press "Ctrl-L"
5.5.2 ) To run the program press "Ctrl-Alt-R"
5.5.3 ) You will see when part of the code run in console area. Errors will show in red. The run is complete when the final line shows ">"

6 ) Running the program from $R$
6.1) Open R Studio
6.2 ) Select "File/Open script" and use the browser to select and open the code file; the code should open in a separate panel. The code is currently called "NEAC_PFA_CL_RR_model_2014"
6.3 ) To set up the code for your PC/laptop, R-click on the code and seroll down to:
\# SET WORKING DIRECTORY ( wd ): In line starting "wd<-" replace the text between the parentheses with the full pathname of the folder containing the code on your laptop (e.g. "D:/Modelling_NEAC/PFA_NCL_R/2014").
\# SET "run_XX": in the lines starting "run $X X$ <-" select which countries you wish to run the assessment for by setting "run- $X X$ ": 1 = run country $X X ; 0=$ do not run. The summaries will only be run if all countries are set to 1 .
\# SET 'PrintFigs': set "PrintFigs" equal to ' 1 ' to output the summary figures (or any other value not to output them).
\# SET 'WinbugsFiles. set "WinbugsFiles" equal to ' 1 ' to output the data files for the Bayesian forecast model (or any other value not to output them).

SET 'PrintCountryTables': set "PrintCountryTables" equal to ' 1 ' to output summary output data for each region that is run (or any other value not to output them).
\# SET 'RunFaroeseCatchSplitEstimation': set "RunFaroeseCatchSplitEstimation" equal to "TRUE" to run the estimation of the Faroes catch composition; otherwise "FALSE" [SEE BELOW]
6.4) You do not need to save your changes before you run the code, but you may wish to save a version to be safe. To do this use "File/Save" or "File/Save As" as normal. It's a good idea to include the extension ".R". NB: You will be prompted to save the file before you close it.
6.5 ) To run the program select "Edit/run all"
6.6 ) You will see when the code runs in the 'R console' panel. Errors will show in red. The run is complete when the final line shows " "
7 ) Running the Faroes stock composition
7.1) The 'Multiannual-data' file contains the latest estimates of the composition of the Faroes catch by European country/regions based on the results of the genetics analysis reported in the 2015 WGNAS report and the 2001-14 PFA outputs. These estimates may be updated if new genetics data are provided or additional years of PFA estimates are to be included.
7.2 ) To run the estimation, ensure that all "Annual-data- XX " files have been updated.
7.3 ) SET 'RunFaroeseCatchSplitEstimation' to "TRUE" and run the model.
7.4 ) The new stock composition parameters will be output in the file "Faroes_split_estimate.txt"; these data should be copied into the end of the "Multiannual-data" file to replace the values already there These data are also required for the Catch Options models.
7.5 ) Reset SET 'RunFaroeseCatchSplitEstimation' to "FALSE" and run the model again to produce full updates of the PFA estimates.
8 ) Output files
The program produces the following outputs (if requested):
8.1 ) National plots: (filenames "Figure-XX")

PDF files showing the national plots currently used in the WG report. This includes: maturing and non-maturing 1SW PFA; returns and spawners for 1SW and MSW; homewater exploitation rates; and total catches (incl. nonreported) for each country ( XX ). It also shows the pseudo stock-recruitment hockey-stock plots for each region; these show the estimated CL, where this is used in the assessment.
8.2 ) Regionat data: (filenames "Region_data_XX_YY")

Excel files showing PFA, returns, catch, exploitation rates and spawners for 1SW and MSW fish and total eggs and lagged egg estimates for each country $X X$ ) and region (YY).
8.3) Input files for Forecast analysis: (filenames:
"Winbugs_Data_XX_YY")
Excel files for each country/region containing mean and s.d. estimates for the simulations for lagged eggs, 1SW returns and MSW returns.
8.4 ) Summary tables by country:

- Median spawner numbers
- Conservation limits and SERs
- Maturing 1SW PFA
- 1SW returns
- 1SW spawners
- Non-maturing 1SW PFA
- MSW returns
- MSW spawners
8.5 ) Summary plot for N-NEAC and S-NEAC
8.6 ) A formatted Excel workbook is set up to link to the output files and format the tables ready for use in the WGNAS report.


## 9 ) Common problems

9.1) The code will crash if an output file (Figure or Table) is left open. The error message (in red) may say:

- Error $\qquad$ : cannot open file 'Fig-XX'
or
- Error in $\qquad$ : cannot open file 'Region_data_XX.csv': Permission denied
9.2 ) It doesn't matter if an input file is open, but the program may not read the latest version if it has not been saved.
9.3 ) More problems to be added .... when they are found!

