# Stock Annex: Salmon (Salmo salar) in Northeast Atlantic and Arctic Ocean

Stock-specific documentation of standard assessment procedures used by ICES.

Stock Atlantic salmon

Working Group on North Atlantic Salmon

(WGNAS)

Created 28 March 2014

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Last updated April 2018

Last updated by

#### General

#### Stock definition

#### **Background**

Atlantic salmon, *Salmo salar* L., have a wide range of life-history strategies. Most forms are anadromous, however, with a juvenile phase in freshwater followed by a period at sea feeding and growing, during which the fish undergo extensive migrations in the open ocean, before they migrate back to freshwater to breed. Most Atlantic salmon return to their river of origin to spawn. This precise homing behaviour has resulted in groups of fish originating in different rivers or tributaries becoming genetically distinct as they adapt to the particular conditions that they face in their home river and along their migration routes. As a result, fish from one river or tributary can differ from fish originating in other rivers/ tributaries which have become adapted to a different set of conditions. These subgroups comprise genetically distinct 'populations' and these are regarded as the basic biological units of the Atlantic salmon species.

Large rivers and their tributaries can support several, genetically distinct populations, each with separate spawning areas within the main-stem of the river or its tributaries. In most instances, however, it is not possible to demarcate clear population boundaries within a river, and managing stocks and fisheries at this level of detail would be very complex. Thus, while there is a need to protect the sustainability of these units, the primary management unit (e.g. for reporting catch statistics and regulating fishing) is generally taken to be the river stock, comprising all fish originating in eggs laid within the river.

Atlantic salmon are native to the temperate and Subarctic regions of the North Atlantic Ocean and there are over 2000 rivers draining into the North Atlantic that support the fish, about 1500 of which discharge into the Northeast Atlantic. In this area, salmon distribution extends from northern Portugal to northern Russia and Iceland, while in the Northwest Atlantic, the species ranges from northeastern USA (Connecticut) to northern Canada (Ungava Bay).

Ideally, the management of all individual river stocks, and the fisheries that exploit them, might be based upon the status of each individual population. This is not always practical, however, particularly where decisions relate to the management of distant

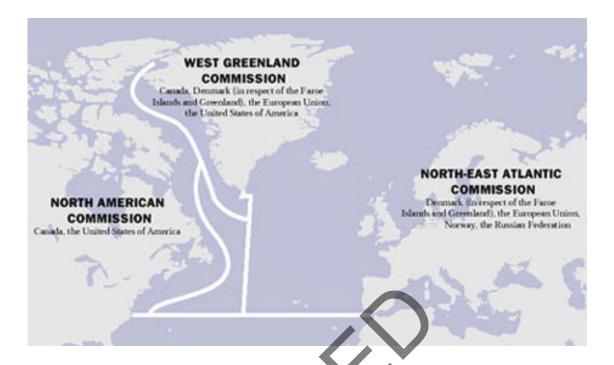
water salmon fisheries, which exploit large numbers of stocks originating in broad geographic areas. WGNAS has therefore had to consider how populations or river stocks should be grouped in providing management advice. For this purpose, groups have been established which fall within the meaning of a stock as 'an exploited or managed unit' (Royce, 1984) and that are consistent with the ICES (1996) definition of salmon 'stocks' as 'units of a size (encompassing one or more populations) which provide a practical basis for the fishery manager'. The issues around the grouping of Atlantic salmon stocks for the provision of management advice are reviewed in detail in Crozier *et al.* (2003). Such stock groupings have typically been referred to as stock complexes.

Salmon mature at various sea ages, typically returning to freshwater to spawn after one to three years at sea, but also sometimes at older sea ages; this varies widely between populations. Those salmon that return after one year at sea are referred to as one-sea-winter (1SW) salmon, or grilse, with older fish categorised as 2SW, 3SW, etc. In practice, however, for management purposes these older sea age fish are typically aggregated and collectively referred to as multi-sea-winter (MSW) salmon. The sea age when salmon become sexually mature depends on genetics as well as growing conditions in the sea, and possibly freshwater, although the precise proximate factors initiating homeward migration are unknown (Hansen and Quinn, 1998). The sea age of Atlantic salmon is important in the context of stock definition since these different groups of fish have different migration routes, return at different times and are differentially exploited in fisheries. Thus, for example, it is only potential MSW salmon that are exploited in the distant water salmon fishery that operates off the west coast of Greenland.

## Management framework for salmon in the North Atlantic

The advice generated by WGNAS is in response to Terms of Reference posed by the North Atlantic Salmon Conservation Organisation (NASCO), pursuant to its role in international management of salmon. NASCO was set up in 1984 by international convention (the Convention for the Conservation of Salmon in the North Atlantic Ocean), with a responsibility for the conservation, restoration, enhancement, and rational management of wild salmon in the North Atlantic. NASCO now has six Parties that are signatories to the Convention, Canada, Denmark (in respect of Faroe Islands and Greenland), the EU (which represents its Member States), Norway, Russia and the USA. While sovereign states retain their role in the regulation of salmon fisheries for salmon originating in their own rivers, fisheries within the jurisdiction of one Party that exploit salmon originating in the rivers of another Party may be regulated by NASCO under the terms of the Convention. This is currently the case for the distant water salmon fisheries at Greenland and Faroes.

NASCO discharges these responsibilities via three Commission areas shown below:



While homewater fisheries are not regulated directly by NASCO, national/ regional jurisdictions seek to comply with NASCO agreements and guidelines in exercising their responsibilities. In particular, NASCO's Agreement on the Adoption of a Precautionary Approach states that an objective for the management of salmon fisheries is to maintain the diversity and abundance of salmon stocks, and NASCO's Standing Committee on the Precautionary Approach interpreted this as being "to maintain both the productive capacity and diversity of salmon stocks" (NASCO, 1998).

NASCO's Action Plan for Application of the Precautionary Approach (NASCO, 1999) provides an interpretation of how this is to be achieved:

- "Management measures should be aimed at maintaining all stocks above their conservation limits by the use of management targets".
- "Socio-economic factors could be taken into account in applying the precautionary approach to fisheries management issues".
- "The precautionary approach is an integrated approach that requires, *inter alia*, that stock rebuilding programmes (including as appropriate, habitat improvements, stock enhancement, and fishery management actions) be developed for stocks that are below conservation limits".

In requesting scientific advice from ICES, NASCO asks for an annual review of events in the salmon fisheries and of the status of salmon stocks around the North Atlantic; NASCO also requests management advice for stocks in each of the Commission Areas. In fulfilling these requirements, three specific purposes have been identified for which stock groupings may be required (Crozier *et al.*, 2003):

- providing descriptions of the status of stocks;
- developing models to estimate and/or forecast pre-fishery abundance (PFA); and
- developing management advice for the distant water fisheries.

Crozier *et al.* (2003) further noted that there is no reason to assume that the same stock groupings should be used for all these purposes, indeed both the criteria used (e.g. geographical or biological features) and the resulting groups are likely to differ.

#### Stock groupings used by WGNAS in providing management advice

As noted above, Atlantic salmon would, ideally, be assessed and managed on the basis of river-specific stock units. In reality, <25% of the rivers with salmon populations in the North Atlantic are so assessed (Chaput, 2012; ICES, 2013). Consequently, stock status is often, of necessity, assessed at broader regional, national and subcontinental scales. While there might be merit in grouping stocks according to biological criteria (which could cross jurisdictional boundaries), it has generally been considered that the difficulties of collecting data in a similar format in different jurisdictions is likely to outweigh the benefits of using such groups (Crozier *et al.*, 2003). It is also recognised that compilations of data on stocks within each jurisdiction are of importance to regional / national managers. As such, regional / national stock groups are typically used by ICES in providing advice on the status of stocks, with additional information compiled on biological groups (e.g. sea ages) as required.

ICES has previously provided information on the status of stocks in the Northeast Atlantic Commission (NEAC) area by region or by country (as well as sea age). For the North American Commission (NAC) area similar information is provided for the USA and the five main provincial regions in eastern Canada: Labrador, Newfoundland, Québec, Gulf and Scotia-Fundy.

In providing management advice for the mixed-stock distant-water fisheries, broader scale stock groupings have been considered appropriate. For the NAC area this is based on the six geographic regions of North America detailed above. For the NEAC area, the following national groupings have been used in recent years to provide NASCO with catch advice or alternative management advice for the distant-water fisheries at West Greenland and Faroes.

Southern NEAC countries	Northern NEAC countries
France	Russia
Ireland	Finland
UK (Northern Ireland)	Norway
UK (England & Wales)	Sweden
UK (Scotland)	Iceland (north/east regions) <sup>1</sup>
Iceland (south/west regions)1	

<sup>&</sup>lt;sup>1</sup> The Iceland stock complex was split into two groups for stock assessment purposes in 2005 (ICES, 2005), largely on the basis of tag-recapture information. Prior to 2005, all regions of Iceland were considered part of the Northern NEAC stock complex.

These groups were deemed appropriate by WGNAS as they fulfilled an agreed set of criteria for defining stock groups for the provision of management advice that were considered in detail at the 2002 WGNAS meeting (ICES, 2002) and re-evaluated at the 2005 WGNAS meeting (ICES, 2005). ICES subsequently noted, however, that provision of catch advice for NEAC stocks in the distant water fisheries should preferably be based on a larger number of smaller management units, similar to those used in the NAC area (ICES, 2010a; 2011). Such an approach was developed at the 2013 WGNAS meeting (ICES, 2013) and indicative catch advice was provided at the country level as

well as the Southern and Northern NEAC stock complexes. ICES (2017) is awaiting feedback from NASCO on the choice of management units.

Salmon from most NEAC stocks mix in the Norwegian Sea in autumn and winter, and are exploited by the fishery at Faroes. While there is evidence that some salmon from NAC rivers have been caught in the Norwegian Sea, they are currently not considered in the NEAC assessments, although this is now under review. Recent genetic information suggests that more North American fish than previously thought were exploited in the fishery at Faroes. Further details on the results of these investigations are provided in Section 3.3.3 of this report and potential options tor accounting for these fish in future catch advice is provided in Section 3.6. To date, consideration of the level of exploitation of national stocks in the Faroes fishery (ICES, 2005) resulted in the proposal that catch advice for the fishery should be based upon all NEAC area stocks and both 1SW and MSW fish.

In contrast, the fishery to the west of Greenland operates in an area where salmon from all North America and some Northeast Atlantic stocks mix in their second summer at sea. Catch advice for this fishery is thus based on non-maturing (potential MSW) fish from all regions of North America, while consideration of the level of exploitation of national stocks in the fishery from NEAC, resulted in catch advice being based upon only Southern NEAC non-maturing 1SW (potential MSW) fish (ICES, 2005).

#### **Fisheries**

Most exploitation of Atlantic salmon is restricted to fisheries close to or within the rivers of origin of the stocks; these homewater fisheries take adult fish that are mainly returning to these rivers to spawn. As noted above, these fisheries are not directly regulated by NASCO since the Parties retain responsibility for the regulation of fisheries for salmon originating in their own rivers. However, NASCO can regulate fisheries undertaken by a Party that take salmon originating in another Party's rivers, such as is the case for the distant-water fisheries at Greenland and Faroes. These fisheries take salmon originating in a large number of rivers over a wide geographical range.

#### The Northern Norwegian Sea Fishery

A longline fishery for salmon in parts of the Norwegian Sea, north of latitude 67°N, commenced in the early 1960s. Several countries participated in this fishery and the pattern of fishing, area of operation and catches changed markedly over the years. At its peak in 1970 this fishery harvested almost 1000 tonnes of salmon.

The Convention for the Conservation of Salmon in the North Atlantic Ocean, which resulted in the formation of NASCO, came into force in October 1983. The Convention created a large protected zone, free of targeted fisheries for Atlantic salmon in most areas beyond 12 nautical miles from the coast. An immediate effect was the cessation of the salmon fishery in the Northern Norwegian Sea outside the Faroes EEZ, with the last catches in this area reported in 1984 (ICES, 2013).

In the late 1980s and early 1990s, NASCO acted through diplomatic initiatives to address fishing for salmon in international waters by vessels registered to non-NASCO Parties. There have been no reports of such activities since this period.

#### The Faroes fishery

The fishery in the Faroes area commenced in 1968 with a small number of vessels fishing up to 70 miles north of the Faroes; initially catches increased slowly up to 40 tonnes in 1977. Danish vessels participated in the fishery between 1978 and 1982 and, at the same time, catches started to increase rapidly, peaking at 1025 tonnes in 1981. Several factors contributed to this increase: the season was extended, more vessels entered the fishery, and the fishery shifted northwards.

From 1982, the Faroese Government agreed to a voluntary quota system, involving a total catch of 750 tonnes in 1982 and 625 tonnes in 1983 ( 255 boats allowed 25 tonnes each). Since NASCO's establishment, regulatory measures or decisions have been agreed by the Northeast Atlantic Commission in most years (Table 1.2.2.1). These have resulted in greatly reduced allowable catches in the Faroese fishery, reflecting declining abundance of the salmon stocks. There has been no commercial salmon fishery targeting salmon around the Faroes since the early 1990s. Catches in the fishery are presented in Figure 1.2.2.1.

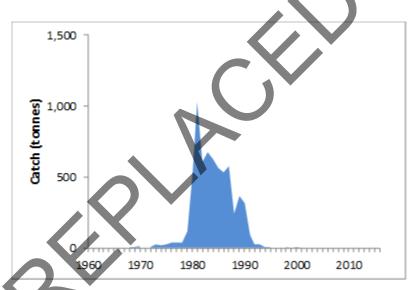


Figure 1.2.2.1. Nominal catch of salmon (tonnes, round fresh weight) in the Faroese longline fishery, 1960–2016.

Table 1.2.2.1. Summary of Regulatory Measures agreed by NASCO for the Faroese Salmon Fishery (courtesy of NASCO).

Year	Allowable catch (tonnes)	Comments/other details in the measures/decisions
1984–1985	625	
1986	-	
1987–1989	1790	Catch in any year not to exceed annual average (597 t) by more than 5%.
1990–1991	1100	Catch in any year not to exceed annual average (550 t) by more than 15%.
1992	550	
1993	550	
1994	550	
1995	550	_
1996	470	No more than 390 tonnes of the quota to be allocated if fishing licences issued.
1997	425	No more than 360 tonnes of the quota to be allocated if fishing licences issued.
1998	380	No more than 330 tonnes of the quota to be allocated if fishing licences issued.
1999	330	No more than 290 tonnes of the quota to be allocated if fishing licences issued.
2000	300	No more than 260 tonnes of the quota to be allocated if fishing licences issued.
2001–2003	No quota set	It is the intention of the Faroese authorities to manage the fishery in a precautionary manner with a view to sustainability, and to make management decisions with due consideration to the advic from ICES concerning status of stocks contributing to the fishery
2004–2006	No quota set	It is the intention of the Faroese authorities to manage the fishery on the basis of the advice from ICES concerning status of stocks contributing to the fishery in a precautionary manner with a view to sustainability and taking into account relevant factors such as socio-economic needs and other fisheries on mixed-stocks.
2007	No quota set	It is the intention of the Faroese authorities to manage any salmo fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2008	No quota set	It is the intention of the Faroese authorities to manage any salmo fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2009	No quota set	It is the intention of the Faroese authorities to manage any salmo fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2010	No quota set	It is the intention of the Faroese authorities to manage any salmo fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.

Year	Allowable catch (tonnes)	Comments/other details in the measures/decisions
2011	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2012	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2013–2015	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs.
2015/2016– 2017/2018	No quota set	It is the intention of the Faroese authorities to manage any salmon fishery on the basis of the advice from ICES regarding the stocks contributing to the Faroese salmon fishery in a precautionary manner and with a view to sustainability, taking into account relevant factors such as socio-economic needs

Note: The quotas for the Faroe Islands detailed above for the period 1984–2000 were agreed as part of effort limitation programmes (limiting the number of licences, season length and maximum number of boat fishing days) together with measures to minimise the capture of fish less than 60 cm in length. The measure for 1984/1985 did not set limits on the number of licences or the number of boat fishing days.

The Faroes salmon fishery operated over winter from November through to May. The salmon caught in the fishery originated almost entirely from European countries with salmon from many countries being present in the area (Jacobsen *et al.*, 2001). Small numbers of tagged fish originating in North America were also recaptured in the fishery (e.g. ICES, 1991), but excluded from catch advice. Genetic investigations, based on salmon scales removed from fish caught in the fishery in the 1980s and 1990s, suggested North American fish may make a larger contribution to the Faroes fishery than originally indicated (ICES, 2015). There was no consistent seasonal trend in the estimated proportion of North American fish in the catches at Faroes and so the overall percentages for 1SW (5.7%) and MSW (20.5%) salmon have been used in subsequent analyses. (ICES, 2015). WGNAS has been asked to consider the implications of the findings in providing future catch advice to NASCO.

The fishery exploited mainly 2SW fish, although some 1SW and 3SW fish were also caught. Small salmon (<60 cm total length) in their first winter at sea were required to be discarded. Large numbers of farmed salmon were also observed at Faroes and there is evidence that farmed salmon escaping from net pens in Norway entered this area (Hansen *et al.*, 1987; Hansen and Jacobsen, 2003). Such farmed fish accounted for a significant proportion of the catch; in the early 1990s, the proportion of farmed fish in this area was estimated at between 25 and 40% (Hansen *et al.*, 1999).

Tagging studies (of adult fish caught in the fishery) have indicated that some fish caught at Faroes were apparently on their way westwards, as they were reported from West Greenland later the same year (Jákupsstovu, 1988). However, salmon tagged at West Greenland were also reported in the area north of the Faroes the following year (ICES, 1984). Thus, salmon of European origin are believed to move through the

Faroese area on their way to the feeding areas in the West Atlantic as well as on their return to homewaters.

#### The Greenland fishery

Limited fishing at West Greenland is reported as far back as the early 1900s, although the present fishery dates from 1959 when local fishermen began setting fixed gillnets from small boats in certain fjords around Maniitsoq (Shearer, 1992). Rapid expansion along the coast followed and from the mid-1960s Faroese and Norwegian fishermen introduced offshore driftnets, followed soon by fishermen from Greenland and Denmark. At around the same time improvements in gear (the introduction of light monofilament nets) enabled fishing in daylight and improved the efficiency of the gear. As a consequence, catches rose quickly reaching a peak of almost 2700 tonnes in 1971. Fishing by non-Greenlandic vessels was phased out in 1972–1975. However, the total catch remained at around 2000 tonnes until 1976 when a TAC of 1190 tonnes was set; the fishery has been regulated since this time. Small catches of salmon are also made on the east coast of Greenland although these are sporadic and restricted by the small number of communities in this area and by drifting polar ice.

Regulatory measures have been agreed by the West Greenland Commission for most of the years since NASCO's establishment (Table 1.2.3.1). These have resulted in greatly reduced allowable catches in the West Greenland fishery, reflecting declining abundance of the contributing salmon stocks. In all but two years since 1998, the fishery has been restricted to an internal-use fishery and commercial export of salmon is not permitted. Catches in the Greenland fishery are presented in Figure 1.2.3.1.

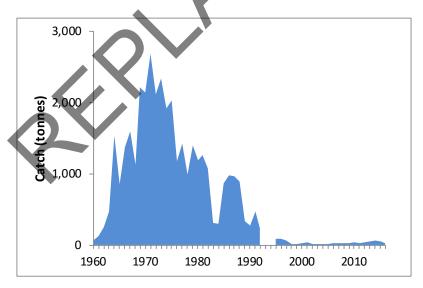


Figure 1.2.3.1. Nominal catch of salmon (tonnes, round fresh weight) in the Greenland salmon fishery, 1960–2016.

The Greenland salmon fishery operates in summer, with a fairly large proportion of the catch commonly being taken in the weeks after the opening of the season in August. Both drift and fixed gillnets continue to operate. The salmon caught in the fishery to the west of Greenland originate in both North America and the Northeast Atlantic. Data on continent of origin in the catch indicate a reasonably even split between fish from North America and Europe in the early 1990s (ICES, 2013). However, the proportion of North American fish in the catch has increased steadily since this time with North American fish comprising 80–90% of the fish caught in recent years.

The salmon caught at West Greenland are almost exclusively fish in their second summer at sea, however, these are non-maturing 1SW salmon destined to return to homewaters as 2SW, or older, fish. Fish from all parts of North America are taken in the fishery, while it is primarily only potential MSW salmon from southern countries in Europe (UK, Ireland and France) that are exploited here. Very few salmon of farmed origin appear in the catches at Greenland, and these are not taken into account in assessments.

Table 1.2.3.1. Summary of Regulatory Measures agreed by NASCO for the West Greenland Salmon Fishery (courtesy of NASCO).

Year	Allowable catch (tonnes)	Comments/other details in the measures	
1984	870		
1985	-	Greenlandic authorities unilaterally established quota of 852 t.	
1986	850	Catch limit adjusted for season commencing after 1 August.	
1987	850	Catch limit adjusted for season commencing after 1 August.	
1988–1990 2520		Annual catch in any year not to exceed annual average (840 t) by more than 10%. Catch limit adjusted for season commencing after 1 August.	
1991	-	Greenlandic authorities unilaterally established quota of 840	
1992	-	No TAC imposed by Greenlandic authorities but if the catch in first 14 days of the season had been higher compared to the previous year a TAC would have been imposed.	
1993	213	An agreement detailing a mechanism for establishing annual quota in each of the years 1993 to 1997 was adopted by the Commission.	
1994	159		
1995	77		
1996		Greenlandic authorities unilaterally established a quota of 174 t.	
1997	57	An addendum to the 1993 Agreement was agreed by the Commission.	
1998	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.	
1999	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.	
2000	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.  A Resolution Regarding the Fishing of Salmon at West Greenland was agreed by the Commission.	
2001	28–200	Under an <i>ad hoc</i> management programme the allowable catch will be determined on the basis of cpue data obtained during the fishery.	
2002	20–55	Under an <i>ad hoc</i> management programme the allowable catch will be determined on the basis of cpue data obtained during the fishery.	
2003–2008	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.	

Year	Allowable catch (tonnes)	Comments/other details in the measures
2009–2011	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
2012–2014	Internal consumption fishery only	Amount for internal consumption in Greenland has been estimated in the past to be 20 t.
2015–2017	Internal consumption fishery only. Greenland unilaterally committed limit the total annual catch for all components of the fishery to 45t in 2015, 2016 and 2017.	The fishery will open no earlier than 1 August and close no later than 31 October each year.  Any overharvest in one year will result in an equal reduction in the catch limit the following year.  Efforts will be made to identify and implement temporal or spatial harvest restrictions that would provide increased protection for weaker stocks.  Greenland will further improve the monitoring, management control and surveillance of its salmon fishery in accordance with the Plan for Implementation and Control Measures in the Salmon Fishery at West Greenland with the objective of achieving full catch accountability.  All Members of the Commission will implement the six tenets.  Greenland will inform NASCO, in a timely manner, of any modifications to the management of the West Greenland salmon fishery, of the outcome of the 2015, 2016 and 2017 fisheries and of progress with the implementation and effectiveness of its Plan for Implementation of Monitoring and Control Measures in the Salmon Fishery at West Greenland. States of origin will explore opportunities to share experiences with Greenland on monitoring, management control and surveillance in the salmon fishery.  A quota of 32 tonnes was set for 2016 fishery. The Commission agreed to review the measure prior to the 2017 fishery.

#### **Ecosystem aspects**

Over the past 20 to 30 years, there has been a marked decline in the abundance of Atlantic salmon across the species' distributional range. Wild Atlantic salmon populations are declining across most of their home range and, in some cases, disappearing (ICES, 2008). Generally, populations on the southern edge of the distribution seem to have suffered the greatest decline (Parrish et al., 1998; Jonsson and Jonsson, 2009; Vøllestad et al., 2009), which may be linked to climatic factors. The decline in salmon abundance has coincided with a variety of environmental changes linked to an increase in greenhouse gases and a corresponding increase in temperatures (IPCC, 2001), which is most likely to have manifest effects at the edge of the species range. However, these areas are often also the ones with higher human population densities and therefore, typically, where potential impacts on the freshwater environment may also be greater. A range of factors in freshwater are known to affect stocks including, for example, contaminants, river obstructions, and changing river flows and temperatures (ICES, 2009b; 2010b; Russell et al., 2012). Such factors have potential implications for the survival of juvenile salmon and their resulting fitness when they migrate to sea as smolts (e.g. Fairchild et al., 2002).

Atlantic salmon occupy three aquatic habitats during their life cycle: freshwater, estuarine and marine. Similar factors contribute to mortality in each of these habitats competition, predation and environmental factors - but despite occurring in different habitats, these are not independent. Conditions experienced within the freshwater environment can affect the survival of emigrating smolts and marine conditions may subsequently modify the spawning success of fish in freshwater.

The decline in salmon populations has occurred despite significant reductions in exploitation, although this does not preclude possible fishery effects. An underlying cause has been a marked increase in the natural mortality of salmon at sea; the proportion of fish surviving between the smolts' seaward migration and their return to freshwater as adult fish (e.g. Peyronnet *et al.*, 2008; Chaput, 2012). For many stocks, return rates are now at the lowest levels in the time-series, even after the closure of marine fisheries. This reduced survival is thought to reflect climatic factors and broad-scale changes in ocean ecosystems as well as factors in freshwater. The exact processes controlling marine survival are relatively poorly understood (Friedland, 1998), although there is growing support for the hypotheses that survival and recruitment is mediated by growth during the post-smolt year, for European stocks at least (Friedland *et al.*, 2009).

Although their habitats are widely separated geographically, there is strong coherence in recruitment patterns between North American and European stock complexes. Recent research suggests recruitment is correlated with ocean temperature variation associated with the Atlantic Multidecadal Oscillation (AMO) (Friedland *et al.*, 2013). It further appears that there are differences in the mechanisms affecting stocks in the Northwest and Northeast Atlantic, with ocean climate variability during the first spring months of post-smolt life most important to the survival of North American stocks, while summer climate variation appears to be more important to adult recruitment variation for European stocks (Friedland *et al.*, 2013). It has been speculated that this may be related to the varying roles of predation pressure and size-related mortality on the two continental stock complexes.

In addition to changes in climate and potential issues operating in both freshwater and marine environments, various other factors have been postulated as possibly contributing to the decline in stock abundance, including predation, aquaculture impacts and the effects of fisheries. Huge increases in aquaculture production of Atlantic salmon over recent decades (see Section 2.2.1 of the WGNAS report) have created some concerns for wild populations. The main potential impacts include: (i) genetic impacts on wild fish; (ii) discharge of organic material and other wastes; (iii) transmission of diseases and parasites (particularly sea lice) to wild populations; and (iv) concerns about obtaining adequate feed resources from an already heavily exploited marine ecosystem. For example, recent investigations in Norway have demonstrated that gene pools of wild salmon populations in a number of rivers have been gradually changed through introgression of genetic material from escaped farmed salmon (Glover et al., 2012; Glover et al., 2013). Sea lice also continue to be regarded as a serious problem for wild salmonids (Skilbrei et al., 2013; Krkošek et al., 2013) affecting their survival and perhaps also their life-history characteristics (Vollset et al., 2014).

As well as declines in abundance, changes in salmon life histories are also widely reported throughout the geographic range of the species, affecting factors such as sea age composition, size at age, age-at-maturity, condition, sex ratio and growth rates (e.g. Nicieza and Braña, 1993; Hutchings and Jones, 1998; Niemelä *et al.*, 2006; Peyronnet *et* 

al., 2007; Aprahamian et al., 2008; Todd et al., 2008). Changes are also manifest in freshwater stages, affecting factors such as the size and growth of parr and the age of smolting (e.g. Davidson and Hazelwood, 2005; Jutila et al., 2006) and run timing (Kennedy and Crozier, 2010; Otero et al., 2013).

#### Data

#### Introduction

Assessment of Atlantic salmon differs from the approaches commonly adopted for other species, for example in respect of the need for at sea surveys and collection of commercial catch per unit of effort (cpue) data. Instead, the assessment of salmon is based mainly on data collected on individual river stocks (e.g. catches and counts of returning fish), which are raised and aggregated to provide estimates of the number of fish returning to homewaters for different stock groupings. These estimates are used, in turn, to estimate abundance at earlier points in the life cycle of the fish and to inform the development of catch advice.

The provision of management advice for the mixed-stock fisheries at Faroes and West Greenland is based on assessments of the status of stocks at broad geographic scales. The North American Commission (NAC) area is divided into six management units, and the Northeast Atlantic (NEAC) Commission area is divided into 19 regions. Assessment of the status of the stocks in these areas is based on estimates of the total abundance - the pre-fishery abundance (PFA) - of different cohorts of salmon at a stage before the distant water fisheries operate. PFA is defined as the cohorts of salmon maturing as 1SW and MSW fish that are alive prior to all the marine fisheries for 1SW salmon (Rago *et al.*, 1993a). The catch advice for the NEAC area is then provided for the northern (N-NEAC) and southern (S-NEAC) stock complexes and for countries.

The models to estimate the PFA of salmon from different areas are typically based on the catch in numbers of one-sea-winter (1SW) and multi-sea-winter (MSW) salmon in each country or region, which are then raised to take account of estimates of non-reported catches and exploitation rates on the two age groups. In some cases, particularly in the NAC area, returns to homewater are estimated by alternative methods, such as counts at fishways and counting fences, or from mark and recapture studies. The estimates of fish numbers returning to homewaters are then raised to take account of the natural mortality (M) between the date that the fish are deemed to recruit to the particular fishery of interest and the midpoint of the timing of the respective national fisheries. A value of 0.03 per month is assumed for M (Section 3.2.3). The date of recruitment of NAC stocks (and thus the PFA date) is taken as August 1st in the second summer at sea because these fish are first exploited in the distant water fishery at West Greenland. However, NEAC stocks recruit to the Faroes fishery during their first sea winter and so PFA is calculated at January 1st (i.e. eight months earlier) for these stocks.

## Input data for assessments-NEAC area

PFA for NEAC stocks is estimated using the run-reconstruction approach described by Potter *et al.* (2004). The model estimates the PFA of both maturing and non-maturing 1SW salmon because both stock components may be caught in the Faroes fishery, and data for both the Faroes and West Greenland fisheries are incorporated into the model.

In order to run the NEAC PFA model, each country provides time-series (beginning in 1971) of catch in numbers, non-reporting rates and exploitation rates for 1SW and MSW

salmon. Best estimates and a measure of the uncertainty or error are provided for the non-reporting and exploitation rate data in order to obtain a measure of the uncertainty in the PFA estimates, since these data are commonly derived from expert opinion. The latest data input variables used in running the NEAC assessment are listed at Appendix 3.

In some instances, the above information has been supplied in two or more regional blocks per country. In these instances, the model output is provided for the regions and is also combined to provide one set of output data for the country as a whole.

The input data for Finland consists solely of catches from the River Tana/Teno. These comprise both Finnish and Norwegian net and rod catches, as the river marks the border between these countries. The Norwegian catches from the river are not included in the input data for Norway.

Where possible, when the input data are themselves derived from other data sources, the raw data are included in the model. This allows the uncertainty in these analyses to be incorporated into the modelling approach. Thus, the catch and sample data used to estimate the catches of Scottish fish in the northeast English coastal fishery are incorporated into the assessments for both UK (England and Wales) and UK (Scotland). For Greenland, catch data are input in the form of harvests (reported and unreported) in weight, along with data from the West Greenland sampling programme.

Descriptions of how the model input data have been derived are presented below for different countries (updated from Crozier *et al.*, 2003; ICES, 2002). The methods used to derive the PFA input data for NEAC countries and options for improving the data are also discussed in Crozier *et al.* (2003).

## Median dates of return to homewater fisheries

NEAC stocks recruit to the Faroes fishery during their first sea winter and so the date of recruitment (and thus the PFA date) is calculated at January 1st. In deriving PFA from the estimates of fish numbers returning to homewaters, it is necessary to take account of natural mortality between the date that the fish recruit to the particular fishery of interest and the midpoint of the timing of the respective national fisheries. The median return date for 1SW and MSW fish for each country/region are provided in the table below. Thus there is about a six to nine month period between the PFA date and the median time of return to homewaters for maturing 1SW fish and 17 to 20 months for non-maturing fish.

Table 2.2.1.1. Midpoint of recruitment to homewater fisheries for NEAC countries/regions.

NEAC Country/ region	1SW	MSW
Northern NEAC		
Russia - Pechora River	8	8
Russia - Archangel / Karelia	7.5	8
Russia - Kola / White Sea	8.5	7.5
Russia - Kola / Barents Sea	7	6.5
Finland	6.5	6
Iceland - north & east	7	6
Norway	8	5
Sweden	8.5	6.5
Southern NEAC		
Iceland - south & west	7	6
UK (Scotland - east)	8	5
UK (Scotland - west)	8	7
UK (N. Ireland - Fo area)		5.5
UK (N. Ireland - FB area)	6.5	6
Ireland	8	5
UK (England & Wales)	8	5
France	8.5	8.5

#### Data inputs for Northern NEAC countries

#### **Finland**

Catch: The catch input to the model of Finland represents an estimate based on catch enquiries and the total number of licences issued. The Norwegian catch from the River Teno has been included in the Finnish catch, which results in a set of input data that effectively represents a single river system. Catch composition is estimated based on catch samples and corresponding scale analyses.

**Level of unreported catch**. Unreported catch is estimated by extrapolating the catches of the fishermen that failed to report their catches, as reporting is not mandatory.

**Exploitation rates**: Exploitation rates in the river fisheries are derived from radio tagging studies in 1992–1993 and 1995, when 70–100 adult fish (1SW and MSW) were tagged yearly in the estuary. Most of the important river fisheries were covered by these experiments.

#### **Norway**

Area split: Salmon catches in Norway are split into four regions on the basis of climatic differences and oceanographic differences among the areas. The areas are: (1) southeast Norway from the Swedish border to the border between Rogaland and Hordaland counties, (2) southwest Norway from the border between Rogaland and Hordaland counties to Stad (3) mid Norway from Stad to Lofoten, and (4) north Norway from Lofoten to the border with Russia.

Catch: Nominal catches of salmon in the four regions are used. In recent years there have been improvements in declaring catches. From 1979 there was a weight split 1SW/MSW (<3 kg/>3 kg). From 1993 the split was changed to 1SW/2SW/3SW (<3 kg/>-7 kg/>7 kg). Mean weight was provided for most groups and used to estimate numbers in the early part of the time-series, but in recent years the reported nominal catch (reported number of killed salmon in river and sea fisheries summed) is being used. In

the input to the PFA model salmon smaller than 3 kg are regarded as 1 SW fish, whereas salmon larger than 3 kg are regarded as MSW fish. The two largest size groups are thus summed into MSW salmon. In the PFA model input the Norwegian catch data for the River Teno have been removed from the Norwegian catches and incorporated in the Finnish catches.

**Unreported catch**: No systematic effort is used to estimate unreported catches. Inputs are guesstimates based on occasional reports from test fishing, surveillance reports, and questionnaires. Currently there is no evidence that the level of unreported catches differs between the four regions. These estimates are provided by the management authorities.

**Exploitation rates**: The rates for the national model are guesstimates. For parts of southeast and southwest Norway they are derived from estimated marine exploitation rates from the River Imsa and the River Drammen. In recent years (from 2009 onwards) exploitation rates for many rivers (>50) have been taken into consideration. These exploitation rates have been obtained using a multitude of methods, mainly from drift counts of spawners or results from counting facilities combined with reported catches in the rivers. The exploitation rates have been adjusted in relation to reduced fishing effort. At present different exploitation levels are used for the different regions, reflecting different harvest regimes in the regions.

For Norway, only data from 1983 onwards have been used for assessment purposes.

#### Russia

Area split: The Atlantic salmon rivers of northwest Russia are split into the following four regions: Kola Peninsula - Barents Sea basin; Kola Peninsula - White Sea basin; Archangelsk Region and the Karelia; and the Pechora River region. The split is based on four regions with separate catch statistics and different biological characteristics of the stocks. For example, the difference in age composition and relative abundance of summer and autumn salmon evident among these four regions has influenced the split.

**Catch:** The declared catch data, in numbers, is available for the full time period (1971 onwards) for all four regions. Catches were allocated to 1SW or MSW age groups on the basis of commercial and scientific catch sampling programmes.

Level of unreported catch: Unreported catches in legal fisheries are estimated from logbooks and catch statistics, by comparing catch survey results with reported catch. Illegal catch is guesstimated and based on local knowledge of fisheries. The major component of the illegal catch in the Barents Sea basin (Kola Peninsula and Pechora River) comes from in-river fisheries and a considerable part of the illegal catch in the White Sea basin (Kola Peninsula and Archangelsk region) comes from coastal areas and this contributes the greatest uncertainties. There is a particular problem with illegal catches on the Pechora River where scientific sampling programmes suggest that the illegal catch on this river is very high. The level of non-reporting increased considerably in the early 1990s due to the economic changes in Russia and temporary reduction of control and enforcement. Since late 2000s the higher level of non-reporting occurred in recreational fisheries due to unclear legislation for reporting. All these factors have been considered in deriving the level of unreported catch for the PFA model.

**Exploitation rates:** Information on exploitation rates is derived from several fisheries in the Kola Peninsula where counting fences are operated and from mark–recapture exercises on the rivers with recreational fisheries. Exploitation rates in Archangelsk

and Pechora are guesstimated. These are the basis of the inputs to the model, regional sea age differences being adjusted on the basis of local knowledge from estimated stock levels.

#### Sweden

Catch: The catch input to the model is based on annual reported commercial salmon catch on the Swedish west coast, and on voluntary reporting from sport fishing in rivers. This reporting is detailed and considered accurate and is handled by the government agency "Swedish Agency for Marine and Water Management" (commercial catches) and the Swedish University of Agricultural Sciences (noncommercial catches). Unfortunately, reporting of catches from non-commercial fishing for salmon with gillnets or rod and line on the coast is lacking. However, due to fishing regulations these catches are small (permits required for trapnets, ban on gillnets in deeper waters, restrictions on the use of gillnets in shallow waters, limited fishing season, large marine protected areas, ban on selling fish, etc.).

There is a large proportion (ca. 60–70%) of reared fish in catches and stocks as a result of compensatory releases of reared smolts (ranching). As all ranched salmon are finclipped the catches of reared fish can be treated separately in the catch statistics. In the reporting from the commercial fishing the catch is not separated into wild and reared fish. The proportion of wild salmon is instead estimated from catch statistics in nearby rivers. Stocking of reared salmon is done in three rivers; all of these also have wild stocks in tributaries.

Catch-and-release is practised in most rivers (only rod-and-line fishing allowed in rivers) but the extent of C&R is not always known. Reported catch is landed fish, but in most rivers a proportion of the fish is released back alive and any subsequent mortality is not accounted for.

Level of unreported catch: Unreported catch, i.e. non-commercial catch of salmon in the coastal area with gillnets and rod and line, is estimated from guesstimates based on expert judgement from regional fishery officers and the Swedish University of Agricultural Sciences. These estimates are supported with catch inventories carried out in 1999 (Thörnqvist, unpublished), 2004 (Swedish Agency for Marine and Water Management), 2008 (Thörnqvist, unpubl.). Generally, the unreported catch is estimated to be 5–10% of the reported catch.

Exploitation rates: Few fish counters are present and tagging data exist mainly for reared stocks, where the fishing pressure is higher than for wild stocks. Input for the PFA model is based on guesstimates. In the index River Ätran, data on size and composition of the spawning run and estimates of exploitation are developed at present. Since 2000, a fish ladder with an automatic counter has provided data on the spawning run in this river. Counter data in combination with results from small-scale tagging in this river are used to provide estimates of exploitation rates. One problem is that exploitation rates differ considerably between rivers. During the period 2000–2014 the average exploitation rates for the Swedish stock as a whole have been estimated at 34% for 1SW and 39% for MSW. The exploitation rate has increased in 2011-2014 due to increased gillnet fishing on the coast. This fishery has been closed and exploitation is expected to decrease in future.

#### **Iceland**

**Area split**: The input data for the PFA model is divided into two areas. Rivers in the west and south of Iceland are combined into one area and rivers in the north and east

into another. This is done on the basis of historic tag recoveries in ocean fisheries (which occurred in different areas) and different climate and oceanic conditions affecting the salmon life cycle, e.g. run-timing, smolt age, and sea age. The southern and western parts of Iceland fall within the NEAC southern area, while the northern and eastern parts of Iceland fall within the NEAC northern area.

Catch: Age-class information is available from individual catch records from logbooks used in the rod fishery. The division into sea age classes is based on a bimodal weight distribution. The 1SW females are <3.5 kg and 2SW females >3.5 kg, while 1SW males are <4 kg and 2SW >4 kg. Scale analyses have shown that the presence of salmon having spent more than two winters at sea and of previous spawners is uncommon and that the categorisation into 1SW and 2SW age classes by weight is regarded as fairly accurate. The net catches are recorded on a daily basis. The age split in the net fishery is derived from the weight distribution in the rod fishery from the same river system or from rivers in the same area.

In the River Ranga in southern Iceland substantial smolt releases have occurred since the early 1990s and have now reached a level of 300 000 to 500 000 smolts annually. Originally, the River Ranga had a small salmon stock with an annual catch of 10 to 90 fish until 1990. The river has very limited habitat for salmon production, but these 'ranched' fish now support a substantial rod fishery. The catch in the River Ranga comprised 23% (18–27%) of the total reported salmon rod catch in Iceland between 2009 and 2013. Since these fish are expected to have very low spawning success in the river they are excluded from the PFA catch input data.

Level of unreported catch: The fishing rights in Icelandic salmon rivers belong to landowners who must, by law, form a fishery association to manage the fishing right. The rod fishing rights are leased to the highest bidder. No ocean or estuary fisheries are allowed. The unreported catch was originally believed to be low with a guesstimate value of 2% applied. With increased use of midwater trawls in pelagic fisheries off the coast of Iceland, new information was provided which suggested an increased level of salmon bycatch. Based on a questionnaire survey, the value of unreported catch was therefore revised after 1995 to a value of 10% of the declared salmon catch. However, more recent analyses of DNA, as well as scale analyses, from salmon sampled as bycatch by Icelandic fishing vessels, indicates a low percentage of Icelandic salmon. Based on this, and other available information, a new estimate of unreported catch is now applied for Iceland at 4% of the declared catch for 1SW and MSW salmon.

**Exploitation rates**: Rates of rod exploitation are based on rivers with fish counters and catch records from logbooks. The estimates of exploitation are 40–50% for 1SW salmon and 50% to over 70% for 2SW salmon. The exploitation estimate for an in-river gillnet fishery is 39% to 52%, with a higher exploitation rate on larger fish. Information on the number of fish subject to catch and release in rod fisheries is also available from logbooks. The proportion of released fish has been increasing since 1996. The reduced exploitation due to catch and release is taken into account in the annual estimate of exploitation for both 1SW and 2SW stock components in the PFA model inputs.

Median return date of 1SW and MSW: Run timing can vary both between years and between areas. The median return date of 1SW and 2SW salmon in south and west Iceland is mid-June and early June respectively. The median date of return is later in the north and east of Iceland, mid-June for MSW and early July for 1SW salmon.

#### **Denmark**

The Working Group collects and routinely reports the annual catch of salmon taken in Denmark. However, the small Danish catches are not included in the assessment process used in developing catch advice for the distant water fisheries.

Catch: The catch input is based on continuously collected reports of salmon taken in the recreational fishery in Danish west coast streams (from Internet sources), which all hold populations of wild salmon. In four of these, where salmon populations have always been found, there is a large proportion of reared (finclipped) salmon in the catch, but these are all F1 offspring from the original populations. In the one catchment in eastern Denmark (Gudenå), where the salmon population is not genetically native to the stream, the annual catch is guesstimated.

**Level of unreported catch**: Unreported catch is expected to be negligible in the western streams because the fishing is closely regulated and controlled by the anglers. In the eastern stream (Gudenå) unreported catch is guesstimated.

**Exploitation rates**: Exploitation rates may be derived from the total catch related to estimates of the total run (calculated by mark–recapture surveys on a three-year cycle in the four streams with original populations on the west coast).

## Data inputs for Southern NEAC countries

#### **France**

Catch: The estimation of salmon catch in France comes from two main sources: (1) mandatory declaration of rod and line catches and from the Adour nets operating in the lower river (scales are sampled from each fish caught) to the Office National de l'Eau et des Milieux Aquatiques (QNEMA), under the Ministère de l'Ecologie, which assumed responsibility from the Conseil Supérieur de la Pêche (CSP) in 2006; and (2) mandatory declaration of catches made by professional net fishermen to Affaires Maritimes, under the Ministere de la Mer, who since 2008 have delegated responsibility for collection and first processing of catch data to the Regional Boards for Sea Fisheries and Aquaculture Catch. At the same time, catches at sea are declared to the Institut Français de Recherches pour l'Exploitation de la Mer (Ifremer), who are responsible for archiving and scientific processing of all fisheries data. Salmon and sea trout catches have not been reliably collated and made available until recently. Since 1985, the 1SW/MSW split has been based on scale interpretation of the in-river catch (based on scale reading) and on a categorisation based on length thresholds for catches in estuaries and at sea. The figures prior to 1985 are not considered as reliable as the later ones.

**Level of unreported catch**: Unreported legal catch for the rod and line fishery has been estimated by catch inquiries made by environmental inspectors of ONEMA on each river. These procedures are still operating in some areas, but estimates are considered less reliable in recent years. The estimation of the professional net fishery catch (Adour Basin) is thought to be reliable and no unreported legal catch is considered to apply.

For most years, the unreported illegal catch is not assessed and a minimal value is provided on a precautionary basis. This unreported illegal catch has been assessed in some years by *ad hoc* inquiries in the estuary of a number of rivers in Brittany (e.g. in 2001) and on the coast (e.g. Baie of Mont Saint-Michel in 2000). The "unreported catch" is included in the nominal catch. No estimates of unreported catch are available for the early part of the time-series (prior to 2001). Thus, the rates input to the model for 1SW

and MSW for the early period are near zero and range from -0.00001 to 0.00001. Higher values in the range 20% to 40% for 1SW and 15% to 30% for MSW fish are applied more recently.

**Exploitation rates**: Exploitation rates are derived from the index River Scorff in Brittany. This is an in-river rate, by rods only, where there are no, or very few, fish thought to be caught on the estuary or coast. Rates are also derived for the Adour river system, where a rough estimation is provided by using the lower values of adult run estimates through facilities in the three rivers flowing to Adour, and the declared catches on the coast, estuary and river, respectively by nets and rods. Some caution is necessary regarding these rates from the Adour given the uncertainties in the different estimates. The rod catch on the index River Nivelle is very small and the probable net exploitation in the estuary and coast is unknown, so exploitation rates are not used for this system. Some data on exploitation rates are also collected by ONEMA on the index River Bresle, but sea trout are the dominant angled species in this river.

#### **Ireland**

Catch: The data are derived from annual declared catches within fisheries districts, management units implemented by Regional Fisheries Boards. Since 2007 river and estuarine specific angling and commercial catch data have been complied. The Fisheries Boards were amalgamated into a single body, Inland Fisheries Ireland, in 2010 which currently takes responsibility for compiling catch statistics. Catches are split by age on the basis of a reported age distribution from 1980 to 1988. In the absence of any other information the mean proportion of 2SW salmon in the series (7.5%) has been used since 1988 and a mean of 10% has been used prior to 1980. Since the introduction of a carcass tagging and logbook scheme for angling and commercial fisheries in Ireland in 2002, sea age classes in the time-series since 2007 have been determined based upon catch dates and weights in accordance with national river stock assessments. The catch is not corrected for returns from releases of smolts for ranching or enhancement but these are not a major component of the catch.

**Level of unreported catch**: The values are guesstimated from local reports and knowledge achieved during catch sampling and fisheries protection activities.

**Exploitation rates:** A coded-wire tagging (CWT) programme has been operated in several rivers in Ireland since 1980. Up to 300 000 hatchery smolts and up to 5000 wild smolts are tagged and released annually. There is also a substantial dataset on wild salmon from the monitored River Burrishoole, providing a further index of wild returns and exploitation rates. Overall, there are estimates of exploitation rates available for three wild stocks and seven hatchery stocks for both 1SW and 2SW salmon. Up to the closure of the marine mixed-stock fishery in 2006, the annual mean of the 1SW wild exploitation index is used as the input data for the lower range of exploitation in the PFA model while the mean of the 1SW hatchery index is used as the upper range. The annual mean of the 2SW wild and hatchery exploitation index was used as the input data for the upper and lower range of exploitation in the PFA model depending on which is higher or lower in that year. Since 2006 the main exploitation input has been from the rod catch which is estimated from coded wire tags estimates for some rivers and also rivers with counters.

#### **UK (England & Wales)**

Catch: Nominal catches for UK (England & Wales) have been derived from the catch returns submitted by netsmen and anglers and split into 1SW and MSW categories

using two different methods. Since 1992, monthly age—weight keys derived from salmon caught in the River Dee trap (an index river) have been used to estimate the age of all rod-caught fish where a weight and date of capture have been provided. This has then been scaled up to the total catch (rods and nets combined) on a pro-rata basis. In earlier years (1971–1991), the age composition of the total catch has been estimated using the mean weight of the fish caught and the mean weight of 1SW and MSW salmon recovered in tagging programmes.

As the contribution of farmed and ranched salmon to the national UK (England & Wales) catch is negligible, the occurrence of such fish is ignored in the assessments of the status of national stocks. However, a large proportion of the fish taken in the northeast coast fishery are destined for Scottish rivers, and these are deducted from the returning stock estimate for UK (England & Wales) and added to the data for UK (Scotland) in the ICES assessment. This proportion is estimated to have declined from 95% of the northeast net catch in the early part of the time-series to 75% in the late 1990s and to around 65% since 2003. This reflects both the steady improvement in the status of the stocks in northeast England and the phase out of the English driftnet fishery since 2003.

Level of unreported catch: All licence holders are required to provide the Environment Agency with details of their catch of salmon and migratory trout and the number of days fished on each river or, for nets, each fishery at the end of the season. Catch returns are received from all net licence holders and from ~90% of full season anglers, and the latter account for the majority of fish caught in a catchment, typically 96–98%. The main correction for underreporting is therefore currently made in respect of perceived inaccuracies in the returns, although more substantial corrections have applied in the past.

There are few independent measures of underreporting in the rod fishery, but these indicate that the level is currently small. A value of 10% is applied for correction purposes based on the method of Small (1991). Historically, underreporting was a much more serious problem. As a result of changes in the licensing and associated catch return system covering UK (England & Wales) in the early 1990s, the percentage of underreporting in the rod catch was estimated to have decreased from ~ 50% to ~20%. Since the mid-1990s, awareness campaigns and enhanced catch reminder systems have further reduced underreporting to the levels currently estimated.

For the net fishery, a figure of 8% has been used in recent years to adjust for the level of underreporting, based on the outcome of surveillance operations. The level may have been substantially higher in the past in certain fisheries, possibly as much as 50%. However, following the successful introduction of logbooks and a carcass tagging scheme in 2009, there is now considered to be minimal underreporting in net fisheries. A figure of 2% has been assumed since 2009.

An earlier questionnaire survey of Environment Agency enforcement staff suggested illegal catches were around 12% of the declared net and rod catch. However, since the introduction of a carcass tagging scheme and a ban on the sale of rod caught fish in 2009, it has been substantially more difficult to dispose of illegally caught fish. Since this time, illegal catches have been estimated to have been reduced to 6% of the declared catch.

**Exploitation rates**: Exploitation rates for a number of monitored fisheries in UK (England & Wales) are derived annually. National exploitation rates have then been estimated by deriving time-series of 'standard fishing units' employed in the salmon fisheries for the period 1971 to the present. For the period 1971 to 1997, these are

calculated from the numbers of licences issued weighted by their relative catching power, which is estimated from historic cpue data; and for the period 1998 to the present, they are calculated from the numbers of days fished by different net categories weighted in the same way. The annual exploitation rates are then estimated by referencing the number of 'standard fishing units' employed over the two periods relative to average age-specific exploitation estimates derived for the 1997 and 1998 seasons.

**Additional information:** Further details on the derivation of estimates within UK (England & Wales) are available in the annual stock status reports (e.g. Cefas, Environment Agency and Natural Resources Wales, 2014), available at:

http://www.cefas.defra.gov.uk/publications/files/SalmonReport2013-final.pdf

#### **UK (Northern Ireland)**

**Area split**: Originally, a single assessment was carried out for UK (Northern Ireland). However, the data used were derived from two fishery management areas (Foyle and Fishery Conservancy Board (FCB) areas), which publish separate catch statistics and have differing fishing regulations. On the basis that stock status in the two areas may differ (Crozier *et al.*, 2003) the two areas were assessed separately from 2001.

Catch: As no commercial fishing has been conducted in the Loughs Agency area since 2010 and in the DCAL area since 2012 the Northern Ireland catch statistics currently (since 2014) rely solely on rod catches. Overall UK (Northern Ireland) rod catch estimates are available since the introduction of a carcass-tagging scheme in 2001. These catch statistics are used as an input in the model. Estimates of sea age composition of the catch for the time-series are based on 1SW/MSW data from adults returning to the River Bush (an index river).

Level of unreported catch: Estimates of unreported catch, as a result of illegal fishing, are based on intelligence reports from DCAL and Loughs Agency fishery officers. These are guesstimates only, with no verification possible. Annual adjustments in unreported catches have been used since tagging programmes started in the mid-1980s. Prior to that, a constant under reporting figure is used, as no annual data are available. The introduction of the carcass tagging scheme in 2001 has led to a reduction in unreported catches.

**Exploitation rates**: Estimates of exploitation rates were historically based on the River Bush microtagging programme. Exploitation from this monitored river (which is in the DCAL fishery area) was used as an input figure for all UK (Northern Ireland) fisheries (Foyle and DCAL areas). However, as currently no commercial fishery for salmon exists in the DCAL and Loughs Agency areas, exploitation rates are based on rod exploitation in the DCAL and Loughs Agency alone.

**Adult counts:** In the DAERA area, counts of adult returns to the rivers Bush and Bann is used as input data for the run-reconstruction model from 2000 onwards. These values are scaled up to the estimated total run using a factor of  $1/(0.67 \pm 0.05)$  for 1SW fish and  $1/(0.61 \pm 0.05)$  for MSW fish, based on more detailed run estimates for 2015. Adult counts are used because these data were more informative about the number of returning adults than very low nominal catches.

**Possible improvements**: A possible improvement would be to have better data available on sea age composition of all Northern Irish fish. Currently the River Bush estimate is applied to all Northern Irish data, but independent data from large river systems like the Bann and Foyle would result in a more reliable country wide estimate.

In addition, a higher return rate for the carcass tagging scheme would result in more reliable estimates of exploitation rates. Recently the carcass tagging return rate in UK (Northern Ireland) has varied between 14% and 55%.

#### **UK (Scotland)**

Area split: The country is divided into eleven statistical regions for the purposes of collating and publishing salmon and sea trout fishery statistics (Marine Scotland Science, 2012). Within the PFA run-reconstruction model, UK (Scotland) is divided into two broad areas (east and west), the split being influenced by the contrasts in topography, river size and the timing and sea age composition of returning fish. The east grouping comprises the East, Northeast, Moray Firth, and North statistical regions, the remaining statistical regions comprise the West grouping in the run-reconstruction model.

Catch: Annual declared catches are collated according to the area split defined above. Reported retained and released catches of wild salmon, taken by net and rod fisheries, are provided separately for two age classes, one sea-winter and multi sea-winter fish. Catch sampling programmes have shown that there is a variable (by region, year, and fishery) proportion of 1SW salmon mis-categorised as MSW salmon in the reported rod catches. The methods used to align abundance estimates at ICES with those used in domestic assessments (see exploitation rates below), also correct for these reporting biases

Level of unreported catch: Previously the unreported catch ranges used in the national model were based on guesstimates made by local managers in some eastern areas of the country (MAFF, 1991). The differences in the ranges used for the east and west groupings were based on a subjective view that unreported catches in the west area were likely to be greater than in the east area due to the demographic and geographic differences between the regions. However, in the absence of empirical evidence with which to support these differences, the current unreported catch rate of 10% was applied throughout the series in both the east and west areas. Error around this value is a subjective estimate of the (normally distributed) uncertainty in the parameter and was set to 5%.

**Exploitation rates**: Abundance is estimated from total rod catch (retained & released) raised using a correction factor to align abundance estimates of home water returns from the run-reconstruction model with those derived in domestic assessment models. Exploitation rates are estimated as all methods retained catch expressed as a proportion of home water returns for both 1SW and MSW salmon.

Additional information: Estimates of spawner abundance take into account estimates of catch and release and natural in-river mortality. Little direct evidence of either source of mortality is available for Scotland. Based on limited information from radiotracking studies, the model assumes catch & release mortality of 10% (Webb, 1998; Smith, Middlemas, and MacLean, 2014) and a further in-river mortality on subsequent escapees of 9% (Milner et al., 2000). Fecundity estimates were also revised to align with domestic assessment methods (Marine Scotland Science, 2017). A time-series of annual points estimates of mean eggs per female for 1SW and MSW salmon were used together with a constant point estimate of sex ratio to estimate egg deposition within the run-reconstruction model. The analysis thus accounts for shifts in fecundity related to observed changes in the lengths of returning fish.

#### Spain

The Working Group collects and routinely reports the annual catch of salmon taken in the recreational rod fisheries in Spain (mainly Asturias). However, the small Spanish catches are not included in the assessment process used in developing catch advice for the distant water fisheries.

#### Data inputs for Faroes and West Greenland fisheries

#### **Faroes**

**Reported catch**: Catches are derived from the landings of salmon caught in the commercial and research fisheries that operated in the Faroes EEZ and the northern Norwegian Sea. Catches for each season (i.e. November in year n to May in year n+1) are assigned to the second year (i.e. year n+1). These fish are classified into 1SW and MSW age groups according to their age (or potential age) on January 1st during the fishery (i.e. a post-smolt caught in November is classified as 1SW).

**Unreported catch**: All fish less than 63 cm total length have been discarded in this fishery and so an unreporting rate of 10–15% (with an error of +/- 5%) has been used for 1SW fish; there is thought to have been negligible non-reporting of MSW fish.

**Catch composition**: Estimates of the proportion of farmed fish in the catch for the period 1981 to 1995 have been derived from scale reading (ICES, 1996; Hansen *et al.*, 1997); prior to 1981 all fish are assumed to have been wild, and since 1997 a value of 0.8 has been used.

Tagged fish originating in North America have been recaptured in the fishery (e.g. ICES, 1991), but excluded from catch advice.

The country of origin of the catch had been estimated based on tagging studies undertaken in the early 1990s (Hansen *et al.*, 1999). These were subsequently replaced by estimates based on genetic analysis.

Genetic investigations, based on salmon scales removed from fish caught in the fishery in the 1980s and 1990s provided an estimated proportion of North American fish in the catches at Faroes. Estimates of 5.7% (1SW) and 20.5% (MSW) have been used in subsequent analyses (ICES, 2015).

The composition of the European component was investigated using individual genetic assignments and gave an overall 1SW stock composition of 84.2% Southern European, 9.0% Northern Europe, 1.2% Icelandic and 5.7% North American (ICES, 2015). The overall composition of the MSW catch was determined as 20.9% Southern European, 58.0% Northern Europe, 0.6% Icelandic and 20.5% North American (ICES, 2015).

It was not possible to use the genetic assignments to estimate the composition of the catches to country/regional level, but they suggested that the composition within the stock complexes was broadly similar to the relative proportions of the PFA estimates and so the breakdown of catches at this level can be made by applying the relative proportions of PFA (ICES, 2015). Sources of uncertainty in these estimates are described in ICES (2015).

#### **West Greenland**

**Catch**: The total nominal catch (i.e. tonnes round fresh weight) in the West Greenland fishery is reported and converted to numbers using a mean weight obtained from the sampling programme.

**Unreported catch**: Estimates of unreported catch were not provided for the period from 1993 to 1999; an annual estimate of non-reported catch, varying from 5 to 20 tonnes was provided by the Greenland representative. Since 2000 a nominal figure of 10 t per year has been provided.

Efforts have been made to provide further information on the level of unreported catch at West Greenland. Since 2002, some assessment of the unreported catch, primarily for commercial landings, has been provided by comparing the weight of salmon seen by samplers involved in the international sampling programme and the corresponding community-specific reported landings. However, since sampling only occurs during a portion of the fishing season, these are considered to be minimum estimates for unreported catch. In addition, there is currently no quantitative approach for estimating the unreported catch for the private fishery. A telephone survey of fishers was carried out following the 2014 to 2016 seasons and provisional findings were provided to WGNAS in 2015 to 2017. These are discussed further in Section 5.1 of the Working Group report (see above); such investigations may provide a basis for revising estimates of unreported catch in future.

Continent of origin: The catch at West Greenland was divided into NAC and NEAC components using scale characteristics until around 2000 and since that time genetic analysis has been used. For the period when scale characteristics were used, the input data to the model are the minimum and maximum estimates of the proportion of NAC fish (from which minimum and maximum proportions of NEAC fish are calculated). For the subsequent period, the inputs are the numbers of NAC and NEAC fish identified in the samples.

## Improvements to NEAC input data

NEAC countries have made ongoing efforts to improve the input data used in assessments. Modifications to input variables are reported by WGNAS in the year in which they are first implemented.

Over recent years, efforts have been made to reduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes). However, the methods used to derive estimates of unreported catch vary markedly between countries. For example, some countries include only illegally caught fish in the unreported catch, while other countries include estimates of unreported catch by legal gear as well as illegal catches in their estimates.

Descriptions of the national approaches used for evaluating unreported catches have been reported at various WGNAS meetings (e.g. ICES, 1996; 2000; 2002; 2010a). In addition, detailed reports describing national procedures for evaluating illegal and unreported catch, and efforts to minimise this, were submitted by parties to NASCO in 2007 in support of a special theme session on this issue. Full details are available at: <a href="http://www.nasco.int/pdf/2007%20papers/CNL(07)26.pdf">http://www.nasco.int/pdf/2007%20papers/CNL(07)26.pdf</a>.

Input data commonly rely on rod catches and the practice of catch and release has become increasingly important in recent years to reduce levels of exploitation on stocks. In the NEAC areas, catch and release estimates from the rod fisheries are not

available from all countries and, when they are, corrections for catch and release mortality are commonly not applied. As the practice of catch and release is increasing, WGNAS have previously recommended (ICES, 2010a) that consideration should be given to incorporating mortality associated with this practice in river-specific, regional and national assessments.

The procedures currently used to incorporate catch and release and unreported catches into regional, national and international assessments are summarised in Appendix 1 (from ICES, 2010a).

One weakness of the NEAC model is that it is heavily dependent upon catch data and estimates of exploitation rate. In most salmon fisheries in the NEAC area, more than half the catch is reported, and in many cases it approaches 100%. However, as stocks have declined, exploitation rates have been reduced to very low levels, and estimates of abundance are therefore becoming increasingly sensitive to this parameter. This inevitably means that uncertainty in the estimates is increasing, and it therefore strengthens the need to make use of alternate sources of information on stock abundance, such as adult counts.

#### Input data for assessments-NAC area

The run-reconstruction model for NAC developed by Rago et al. (1993a) is used to estimate the PFA of non-maturing 1SW salmon of North American origin (beginning in 1971). Only the West Greenland fishery is of relevance in the context of distant water exploitation of NAC stocks. This fishery exploits predominantly (>95%) 1SW non-maturing salmon (destined to return primarily as 2SW salmon) and hence it is only necessary to estimate the abundance of this age group prior to the fishery at Greenland. The other fish taken in the fishery represent 2SW and older non-maturing salmon and previous spawners (ICES, 2003). However, PFA estimates for maturing 1SW salmon as well as large salmon (containing all MSW age groups of salmon including repeat spawners) are derived from the run-reconstruction model.

The starting point for the reconstruction requires estimation of the returns of 2SW salmon, small salmon and large salmon to the six regions in eastern North America: Labrador, Newfoundland, Québec, Gulf, Scotia-Fundy, and USA. With the progressive closure of commercial fisheries (1984 for the Gulf and Scotia-Fundy regions; 1992 for Newfoundland; and 1998–2000 for Labrador and Québec) abundance estimates have relied less on harvests and increasingly on estimated returns to rivers raised to production areas. The returns for each region are estimated with the uncertainty defined by a range of minimum and maximum values based on the best information available for each region (Chaput *et al.*, 2005).

The annual pre-fishery abundance of non-maturing 1SW fish for year i, destined to be 2SW returns (excluding 3SW and previous spawners), represents the estimated number of salmon at West Greenland prior to the start of the fishery on August 1st. Definitions of the input variables used in the model are given in Table 2.3.1. The PFA estimate is constructed by summing 2SW returns in year i+1 [NR2(i+1)], 2SW salmon catches in commercial and Aboriginal peoples' food fisheries in Canada [NC2(i+1)], and catches in year i from fisheries on non-maturing 1SW salmon in Canada [NC1(i)] and Greenland [NG1(i)].

Table 2.3.1. Definitions of key variables used in continental run-reconstruction models for North American salmon.

i	Index for PFA year corresponding to the year of the fishery on 1SW salmon in Greenland and Canada $$
М	Natural mortality rate (0.03 per month)
t1	Time between the midpoint of the Canadian fishery and return to river = 1 month
S1	Survival of 1SW salmon between the homewater fishery and return to river {exp-M * t1}
H_s(i)	Number of "Small" salmon caught in Canada in year i; fish <2.7 kg
H_I(i)	Number of "Large" salmon caught in Canada in year i; fish >=2.7 kg
AH_s	Aboriginal and resident food harvests of small salmon in northern Labrador
AH_I	Aboriginal and resident food harvest of large salmon in northern Labrador
f_imm	Fraction of 1SW salmon that are immature, i.e. non-maturing: range = 0.1 to 0.2
af_imm	Fraction of 1SW salmon that are immature in native and resident food fisheries in N Lab
q	Fraction of 1SW salmon present in the large size market category; range = 0.1 to 0.3
MC1(i)	Harvest of maturing 1SW salmon in Newfoundland and Labrador in year i
i+1	Year of fishery on 2SW salmon in Canada
MR1(i)	Return estimates of maturing 1SW salmon in Atlantic Canada in year i
NN1(i)	Pre-fishery abundance (PFA) of non-maturing 1SW + maturing 2SW salmon in year i
NR(i)	Return estimates of non-maturing + maturing 2SW salmon in year i
NR2(i+1)	Return estimates of maturing 2SW salmon in Canada
NC1(i)	Harvest of non-maturing 1SW salmon in Nfld + Labrador in year i
NC2(i+1)	Harvest of maturing 2SW salmon in Canada
NG(i)	Catch of 1SW North American origin salmon at Greenland
T2	Time between the start of the fishery at West Greenland (August 1) and return to the coast of North America = 10 months
S2	Survival of 2SW salmon between August 1 (at West Greenland) and return to the coast of North America {exp-M * 12}
MN1(i)	Pre-fishery abundance of maturing 1SW salmon in year i

## Data inputs for NAC

The latest data input variables used in running the NAC assessment are listed at Appendix 4. More detailed descriptions of how the model input data have been derived for each region of North America are presented below.

#### Labrador

For Labrador stocks, it was thought to be inappropriate to develop total recruits from angling catches and exploitation rates similar to techniques used for rivers in insular Newfoundland. The problem with using angling catches to derive returns for Labrador is, that until 1994, there were no estimates of exploitation rates available other than for the salmon population of Sand Hill River and these were 20 years out of date. Also, because Labrador coastal rivers are isolated, the exploitation rates are low and highly variable depending on the presence of an angling camp and its success in attracting guests as well as the nearness of local communities. Thus, exploitation rates would depend, and vary from one year to the next, on the success of angling camps in attracting anglers and may not be applicable to other Labrador rivers. Thus, all estimates of returns and spawners until 1998 were based on commercial catches as the only source available of usable continuous time-series of data.

#### Before 1998

The general approach is to use exploitation rates to convert commercial catches of small and large salmon in Labrador to total population prior to the commercial fishery. River returns and spawners were estimated by subtracting the commercial catch from these populations, and accounting for non-Labrador interceptions. The estimated number of Labrador origin large returns is calculated as:

$$LR = (CC*PL) / u$$
 (1)

where,

LR = Labrador returns, PL = proportion Labrador origin, CC = commercial catch, and u = exploitation rate

The estimated number of Labrador origin small returns is determined from equation (1) but using commercial catches of small salmon.

Parameter values for sea age and the proportion of salmon of Labrador origin comes from the sampling program in the commercial fishery, 1974–1991. In 1997, commercial sampling resumed with samples being collected throughout the fishery at Makkovik and Rigolet in SFA 1 and Cartwright and St Lewis/Fox Harbour in SFA 2. River age distribution of commercial samples of small and large salmon from Labrador have been found to consist, on average, of about 75-80% river age 4 and older in SFAs 1 & 2. The commercial samples came from commercial catches sampled in Labrador at several sites along the Labrador coast including Square Islands (SFA 2) and at Nain (SFA 1) (Anon, 1993b). In total, 46 320 salmon were sampled for scales and aged. Labrador salmon stocks are thought to contribute about 70% of the total production of four year, and older, river age salmon, with the other 30% coming from northern Québec. Thus, when non-Labrador salmon are factored in at 30% applied to the river age distribution, then 60-80% of the harvest of small and large salmon (PL) in Labrador are of Labrador origin (Anon, 1993b). In 1997, in SFA 1, the percentage of the commercial catch that was of Labrador origin was for large salmon 68% (95% C. I. 64.3-72.5%); whereas for small salmon it was 39% (95% C.I. 35.6%-41.6%). In 1997, in SFA 2, the percentage of the commercial catch that was Labrador origin was for large salmon 92% (95% C. I. 88.4-95.2%); whereas for small salmon it was 80% (95% C.I. 74.8%-85.0%).

Exploitation rates (u) were calculated from the smolt tagging study in 1969–1973 on Sand Hill River (Reddin, 1981; Reddin and Dempson, 1989). Exploitation rates of 0.28 to 0.51 for small salmon and 0.83 to 0.97 for large salmon from the tagging study were changed to base exploitation rates of 0.3 to 0.5 on small salmon and 0.7 to 0.9 on large salmon and were assumed to apply to all of the salmon populations in SFAs 1, 2, and 14B for the period of 1969–1991 (Anon., 1993b). After 1991, due to the Management Plans for the commercial fishery in Labrador and Newfoundland, several changes occurred that would reduce exploitation of Labrador origin salmon. These changes include: (1) reductions in effort as commercial salmon fishermen chose to sell their licences from a buy-out agreement begun in 1992, (2) a moratorium on commercial fishing in Newfoundland would increase the number of Labrador salmon in Labrador coastal waters, and (3) season reductions due to the varying opening dates and early closures from the quotas applied in 1995 and 1996. The effects of these changes were quantified in the exploitation model as follows:

$$u=1-e^{-aF}$$
 (2)

where: a = fraction of the 1991 licensed effort remaining in 1992–1996.

In 1994–1996, the licensed effort for all of Labrador was 37% of the 1991 level of 570 licences, in 1993 it was 55%, and in 1992 it was 87%. In any given year, it was assumed that 90% of licensed fishermen were active. Fishermen reported during public consultations that in 1995 and 1996 many licensed salmon fishermen did not fish for salmon in 1995–1996 but fished for crab instead. This was verified by Fisheries Officers who reported that of the 218 licensed salmon fishermen only 132 were active in 1996. Another method of obtaining actual effort information is also available since, beginning in 1993 commercial fishing vessel (CFV) numbers have been recorded on sales receipts issued to fishermen by fish plants. Enumeration of licensed salmon fishermen actively fishing was made by determining the number of CFVs in the Statistics Branch catch records. Active effort in 1991 and 1992 was assumed to be 90% as it was in 1993 and 1994 from the CFV file. Thus, the exploitation rates (u) were modified due to effort reductions in equation (2) using estimated active licences from 1991 as a base and the number of active licences in 1995, 1996 and 1997. The modified exploitation rates (ue) for 1992–1997 used the licensed effort in equation (2).

The tagging study on Sand Hill River, 1969–1973 showed that Labrador small and large salmon were not only caught in Labrador, but also in the commercial fisheries along the northeast coast of Newfoundland (both small and large) and at West Greenland (large only) (Anderson, 1985). For small salmon, out of a total of 100 (1SW) tag returns there were 24 from Newfoundland. For large salmon, out of a total of 137 (2SW) tag returns there were 41 from Newfoundland.

For 1992–1997: the moratorium on commercial fishing in Newfoundland would have released small and large salmon to Labrador. The effect of salmon released from Newfoundland in 1992–1996 was evaluated against the exploitation rates as follows:

un = 
$$(1-((24*(1-ue))/100))*ue$$
, for small salmon, and  
un =  $(1-((41*(1-ue))/137))*ue$ , for large salmon (3)

The new estimates of fishing mortality (un) in 1992-1994 included adjustments for the closure of the commercial fishery in Newfoundland based on the results of the Sand Hill River tagging study. Season reductions due to the varying opening dates and early closures from the quotas applied in 1995 and 1996. In 1995, adjustments were made to account for the new opening date for the commercial fishery in Labrador of July 3 changed from June 20 the previous year. For 1995, the accumulative effect of these, weighted to SFA catches, was to reduce the catch so that for small salmon the current catch represents 86.0% of small salmon and 62.7% of large salmon. In 1996, the opening date reverted to June 20 but the quota levels resulted in early closures in SFA 2 of 2A -July 10, 2B - July 8, and 2C - July 2 while SFA 1 and 14B did not close. For 1996, the accumulative effect of these weighted to SFA catches was to reduce the catch so that for small salmon the current catch represents 53% of small salmon and 61% of large salmon. In 1997, the opening date remained at June 20 but the quota levels resulted in early closures in SFA 2 of 2A -July 12, 2B - July 15, and 2C - July 13 while SFA 1 closed on October 15 as the quota was not caught. For 1997, the accumulative effect of these early closures was to reduce the catch so that for small salmon the current catch represents 47% of small salmon and 64% of large salmon. The season changes reduce catches and hence lower exploitation rates. The effect of shorter seasons in 1995, 1996 and 1997 was evaluated against the exploitation rates in section B as follows:

$$US = UN * SC$$
, for small salmon, where SC is season change, and  $US = UN * SC$ , for large salmon (4)

The new estimates of fishing mortality including effort reductions, adjustments for the closure of the commercial fishery in Newfoundland, and shorter seasons due to opening dates and quotas results in the following exploitation rates which were applied to catches. The cumulative effect of factors A, B, and C is to reduce exploitation on Labrador origin salmon.

Labrador origin 2SW returns (LR2SW) were derived from eq. 1 by:

$$LR2SW = LR * P2SW$$
 (5)

where: P2SW = proportion of the large salmon that is 2SW salmon.

The SR1SW were calculated as in equation (5) but using P1SW which is the proportion of the catch that is 1-sea winter in age and maturing to enter freshwater and spawning in the year of capture. The parameter values for P1SW of 0.1 to 0.2 come from Anon. (1991).

The 2SW component was estimated separately for salmon caught in SFA 1, 2 and 14B. In SFA 1, commercial sampling at Nain of large salmon showed the proportion of 2SW were on average about 84% (n=6542), 1977–1991. Thus, a range of 0.7–0.9 was used for SFA 1. In SFA 2, commercial sampling of large salmon averaged 69% (n=4793) 2SW salmon, 1977–1991. There were no commercial samples available for SFA 14B. Thus, for SFAs 2 & 14B a range of 0.6–0.8 was used. For the 1SW component, commercial samples at Nain in SFA 1 of small salmon showed the proportion of 1SW salmon were on average about 94% (n=4757). In SFA 2 the 1SW component was on average about 97% (n=8872) of small salmon. There were no samples from commercial sampling in SFA 14B. In 1997, aged commercial samples indicated that the previous range was acceptable.

Total river returns of 2SW salmon (TRR) were calculated as follows:

$$TRR = LR2SW / (1-us)$$
 (6)

The total river returns of small salmon are also calculated by equation 6 but from SR. Spawning escapement (SE) or spawners was calculated according to the formula:

$$SE = TRR - AC, (7)$$

where:

AC = angling catch which includes retained catch plus 10% of hook & released mortality for released salmon.

A couple of modifications were made to the estimation procedure for Labrador in 1997. First, determination of exploitation rates were calculated separately for SFA 1, 2 and 14B using the active effort individually for each SFA. For SFA 1, the active number of licences declined from 141 in 1991 to 39 in 1997. For SFA 2, the active number of licences declined from 320 in 1991 to 99 in 1997. For SFA 14B, active licences declined from 52 in 1991 to 0 in 1997 when the fishery was closed. Exploitation rates determined as in equations 2, 3 and 4 are: SFA 1 - small was 0.0735 to 0.1399 and - large was 0.2221 to 0.3959; and SFA 2 - small was 0.0384 to 0.0728 and - large was 0.1589 to 0.2799.

Numbers of small and large salmon for SFAs 1 & 2 were estimated from the exploitation model while for SFA 14B the results of assessments on Forteau Brook and

Pinware River were expanded to include all the watersheds in SFA 14B. Returns to SFA 14B were 663 to 1545 small salmon and 146 to 327 large salmon.

Total mortalities of small and large salmon were accounted for by summing commercial catches of small salmon in Labrador and Newfoundland, large salmon in Labrador, Newfoundland, and Greenland, angling catches in Labrador of small and large salmon including 10% of the hook and released salmon, and small and large spawners. All of the above mortality estimates except catches of Labrador salmon in Newfoundland, 1969–1991 and Greenland could be obtained from equations 1 to 7. Catches in Newfoundland and Greenland were assessed as follows:

Greenland: for 1969–1992 and 1995–2004, removals of Labrador salmon by the Greenland fishery were assessed from data based on the sampling program in commercial fish plants at West Greenland (Anon, 1996). The Greenland fishery catches salmon that would have returned to homewaters as large salmon in the year following the Greenland fishery. Numbers of Labrador salmon were determined by converting catches in kg to numbers of salmon of 1SW North American origin that were of river age 4 and older. The number of Labrador salmon were estimated by assuming that 70% of the production of 4-year and older river age salmon are from Labrador (Anon, 1993b).

**Newfoundland**: for 1969–1991, catches of Labrador small and large salmon in Newfoundland were included in total mortalities as the product of the ratio of tags caught in Newfoundland to Labrador and the catch in Labrador. For small salmon the ratio was (24/(100-24)) = 0.32 and for large salmon it was (41/(137-41)) = 0.43.

#### 1998-2001

For the years, 1998–2001 when only one or two counting projects took place in Labrador, the raising factors previously used and explained of 1.04 to 1.49 for small salmon and 1.05 to 1.27 for large salmon were used to estimate returns and spawners for Labrador from the overall PFA minus catches in Greenland, as was the case in previous years. However, in this case returns to rivers were derived for Labrador by subtracting landings in food fisheries. Also, catches in 1994–2006 were updated to reflect changes made to catch statistics in Labrador from the Licence Stub Return System. Procedures for the collection and compilation of commercial and angling fishery data are described in Ash and O'Connell (1987) for fishery years 1974–1996. For years 1969–1974, commercial catch data came from Anon. (1978). In 1997, the angling catch statistics were converted to a licence stub system (O'Connell *et al.*, 1998) which continues to present.

#### 2002-present

Counting projects occur on three to four Labrador rivers; out of about 100 extant salmon rivers. Because they occur on the same rivers each year, it is possible to extrapolate from abundance for small and large salmon per accessible drainage areas in these monitored rivers to unsurveyed ones in the remainder of Labrador. The accessible drainage areas were 9267 km² for Lake Melville (SFA 1A), 25 485 km² for Northern Labrador (SFA 1B), 28 160 km² for Southern Labrador (SFA 2), and 2651 km² for the Straits Area (SFA 14B). Accessible drainage area in the counting facility rivers was 1878 km² resulting in an expansion factor of 35 to one. Not all rivers in Lake Melville were included due to a lack of information on presence of salmon populations in rivers in this region of Labrador. Lake Melville rivers whose drainage areas were

included are Sebaskachu, Cape Caribou, Goose, MacKenzie, Kenamu, Caroline and Traverspine.

Abundances for SFAs 1A and 1B were derived from English River returns with maximum and minimum values developed using the observed variability of relative abundances in SFA 2. Total returns and spawners for Labrador are estimated by Monte Carlo simulation based on 10 000 random draws from the range of values assuming abundances per km² of accessible drainage were uniformly distributed. The relative abundances (per km²) for each SFA were then multiplied by the total accessible drainage area to derive total returns of small and large salmon. Ranges of values were developed to convert numbers of small and large salmon to numbers of 1SW and 2SW salmon from scale age information collected from counting fences and angling fisheries in Labrador. A bootstrap procedure was used to develop estimates of the proportions of sea age 1 salmon in estimates of large salmon returns and spawners, proportions of sea age 1 salmon in the estimates of large salmon returns.

Sea age correction factors were:

Small to 1SW - 96 to 100% Large to 2SW - 60 to 71% Small overlap in large - 12 to 21%

Spawners of 1SW and 2SW salmon were derived by subtraction of angling catches including an estimate of hook and release mortalities (10%) from the returns.

#### Newfoundland

Inputs for the run-reconstruction model for Newfoundland include estimates of small, large and 2SW returns and spawners to rivers (minimum and maximum). The methods used to estimate returns and spawners to the rivers in Newfoundland are described by Reddin and Veinott (2010). In brief, returns and spawner estimates were derived from recreational fishery exploitation rates of retained small salmon for rivers with enumeration facilities; and ratios of large to small salmon were utilized to estimate large salmon. Exploitation rates were then applied to all rivers with reported angling catches. A non-parametric bootstrap technique was used, whereby exploitation rates and ratios of large to small salmon from rivers with enumeration facilities were chosen at random with replacement. The 95th confidence interval from 500 iterations of the weighted exploitation rate and ratio of large to small salmon was applied to angling catches on a Salmon Fishing Area (SFA) basis. The midpoint of the 95th confidence interval was used as the minimum and maximum estimate returns of large and small salmon in each SFA. Estimates of 2SW returns are based on the expected proportion of 2SW in the large salmon category (≥63 cm). Commercial and recreational angling catches were derived as described for Labrador (2.3.1.1). Spawners in all years were determined as the returns to rivers minus angling catches including an adjustment for hook-and-release mortality.

#### Québec

In order to estimate abundance of stocks, rivers were classified into six categories (C1–C6) depending on the information available to estimate salmon returns (according to the method of Caron and Fontaine, 1999), with C1 being the most reliable evaluation and C6 the least. C1 corresponds to a river where the evaluation of the returns is based on a counting method, either from a fence or from a visual count through snorkelling

or from a canoe. C2 uses the same evaluation, but without knowing the number of small and large salmon, which is then estimated from proportions reported in the sport fishing landings and, if necessary, the catch and release. Salmon returns on C3 rivers are determined based on multiple correlation factors, using catch number, fishing effort, season duration and river accessibility distance (Guillouët, 1993).

When estimation of the returns using a C1–C3 category is not possible, and when data of returns from previous years are available, the C4 category is used. C4 assumes that interannual variations in salmon returns in the targeted river are approximately the same as variations observed in the other rivers of the corresponding region. Category C5 is for rivers where only landings data are available. In these rivers the salmon run is estimated from the average regional exploitation rate. Finally, a few small rivers have essentially no available data. C6 then assumes that the run is related to the available river salmon habitat and is estimated with respect to rivers of the same area for which run estimates and salmon habitat area are known. Estimated numbers of returns from C4 to C6 cannot be used to assess relative to attainment of conservation limits. However, they provide at least approximate numbers to estimate returns and spawners for salmon rivers in Québec.

The evaluation of the uncertainty associated with return estimates depends on the river category. For C1 and C2 rivers, the correction factor for the minimum and maximum number of returns is +5% and +10% for all rivers with a fish ladder and for all others in zones Q1 to Q3 and Q10. The correction factor for rivers with darker water from zones Q5, Q6 and Q7 is rather +10% and +30%. For the other categories, an uncertainty of ±25% is associated with salmon return estimates, except for category C3 where calculation depends on the method of Gutllouët (1993).

The number of spawners is obtained using the return estimate minus all river catches, which include landings and other types of removal. In most cases, river catches include landings from sport fishing only, which may be conducted by native people such as that on the Betsiamites River. The other types of removal are of limited number and include mainly natural mortality, salmon captured for hatchery use and subsistence fishing when practised in river.

Overall return estimates for all Québec rivers are obtained by adding in-river salmon returns, commercial fishing (when operated), native people subsistence fishing when practised in estuaries and an estimate of non-registered landings. However, little scientific data are available on non-registered landings and thus, estimates are based on good judgment, following consultations with regional biologists.

#### Gulf

Estimation of returns and spawners are developed for the four salmon fishing areas of Gulf Region (SFAs 15 to 18).

#### **SFA 15**

The major river in this area is the Restigouche River. The returns and spawners are estimated for the Restigouche River exclusive of returns to the Matapedia River, which are included in Québec zone Q1. The Restigouche River stock assessment is based on angling catch with assumed exploitation rates between 30% (min.) and 50% (max) with estuary catches added back after the estimates of returns. Return and spawner estimates for SFA 15 are based on Restigouche River data, scaled up for SFA 15 using angling data. The return and spawner estimates for SFA 15 are derived from the return and spawner estimates for Restigouche (New Brunswick). The minimum and

maximum return and spawner estimates are derived from the minimum and maximum ratios of angling catch in all of SFA15 relative to angling catch in Restigouche (New Brunswick) (min = 1.117; max = 1.465). Harvests represent retained angling catch plus 6% catch and release mortality for released fish. The proportion of 2SW in large salmon numbers is based on aged scale samples from angling, trapnets, and broodstock. In the years when no scale samples analysis is available, a mean value of 0.65 is used.

#### **SFA 16**

The most important Atlantic salmon river in SFA 16 is the Miramichi River. The Miramichi makes up 91% of total rearing area of SFA 16 and returns to the river are assessed annually. For 1971 to 1991, minimum and maximum values are based on capture efficiencies of the Millbank estuary trapnet representing a lower CI of -20% of the estimate and upper CI of 33% of the estimate. For 1992 to 1997, minimum and maximum are lower and upper CI and based on estimate bounds of -18.5% to +18.5%. Since 1998 to the present, minimum and maximum are 5th and 95th percentile range from a Bayesian hierarchical model used in the assessment. Returns to SFA 16 are Miramichi returns (Minimum, Maximum) / 0.91. Proportion 1SW in small salmon is from scale ageing; proportions have varied from 0.97 to 1.0. Proportion 2SW in the large salmon category is obtained from scale ageing. Spawners are returns minus harvests. For 1998 to 2017, the harvest of large salmon is estimated as the sum of the aboriginal fisheries harvests for large salmon and 1% of the large salmon catch (30% exploitation rate, 3% catch and release mortality). Prior to 1995, the harvest of small salmon is estimated as 30% of the small salmon return plus the harvest from the aboriginal fisheries. During 2015 to 2017, when mandatory catch and release management measures were in effect, the fisheries related losses (harvests) of small salmon were estimated as the sum of the aboriginal fisheries harvests for snall salmon and 1% of the snall salmon catch (30% exploitation rate, 3% catch and release mortality).

#### **SFA 17**

For 1970-1994, small returns are estimated from retained small salmon catch in the Morell River divided by the river-specific exploitation rate. Salmon catch in the Morell River was estimated in 1970-1990 by DFO Fisheries Officers; and in 1991, 1992, and 1994 by angler mail-out surveys. The number of small retained salmon in 1993 was not recorded, so the number used is the mean for 1986–1992. For 1970–1993, exploitation rate was taken as the mean of exploitation rates estimated for 1994, 1995, and 1996 (0.317). For 1994, exploitation rate was 0.34. The min and max of small returns are calculated using exploitation +/- 0.1; e.g. 0.34 +/- 0.1 gives 0.24 and 0.44. Large returns = (number of small returns/proportion small) - number of small returns. For 1970–1980, proportion small is calculated from numbers of small and large salmon in the angling catch of each year. For 1981-1994, proportion small is taken from counts at the Leards Pond trap on the Morell River. Small spawners = number of small recruits-number of small retained. Large spawners = number of large recruits - number of large retained. In 2012, the Province of Prince Edward Island discontinued the sale of recreational fishing licences for Atlantic salmon. Instead, anglers who purchased a trout licence are authorized to also fish Atlantic salmon. Since it was no longer possible to assemble a list of salmon anglers, the salmon angler survey was discontinued starting in 2012. In the absence of salmon angling data for 2012 and subsequently, catch statistics estimated for 2011 are used for 2012 and subsequent years.

Spawner estimates for 1995 to the present are derived from redd counts in 23 rivers. For years and rivers in which redd counts are unavailable, redd numbers are estimated by linear interpolation from the preceding and succeeding count year. Redd numbers in years prior to the first count are taken as the first count. Redd numbers in years after the last count are taken as the last count. Female spawners are estimated from the ratio of 3.357 redds/female spawner, measured in the West River in 1990. Total spawners are estimated from size-specific sex ratios derived from counts at Leards and Mooneys Ponds, Morell River, in 1986–2001. The proportion of salmon that are large is assumed to be 0.5 in the Cains, Carruthers, Trout (Coleman), Morell, Cardigan, West, and Dunk Rivers, and 0.9 in all other rivers. Spawners are presented as Min (estimated spawners -20%) and Max (estimated spawners + 20%). Returns are spawners + total estimated fishing mortality, including angler catches, hook-and-release mortality, and native harvests. Angler catches and hook-and-release mortality are estimated from angler card surveys. Returns are presented as Min (estimated returns -20%) and Max (estimated returns + 20%). It is assumed that large salmon and 2SW salmon are equivalent.

#### **SFA 18**

Returns and spawners to SFA 18 are derived from estimates of returns and spawners to the Margaree River, adjusted for the ratio of the SFA 18 angling catch to the Margaree River catch. For small salmon, the ration of SFA 18 catch to Margaree catch varies between 1.15 and 2.71 for years 1984 to 2004. For large salmon, the ratio of SFA 18 catch to Margaree catch varies between 1.08 and 2.32 for years 1984 to 2004. Returns to Margaree River are estimated using various techniques.

- 1970 to 1983 angling catch divided by range of exploitation rates with maximum exploitation rate of 0.27 and minimum exploitation rate of 0.215;
- 1984 to 1986 based on annual assessments;
- 1987 to present angling catch and effort data from logbooks and provincial licence stubs are used to derive the returns. The catchability coefficient per rod day is estimated from angling catch and effort data for the years 1988 to 1996 when mark and recapture programmes were used to estimate returns, independently from angling data.

Spawners for 1970–1983 equal returns minus removals. Spawners for 1984 to the present equal returns minus catch for small salmon and returns minus catch, corrected for 5% mortality, for large salmon. 2SW salmon represent between 0.77 and 0.87 of large salmon returns and spawners.

#### Scotia-Fundy

Salmon originating in rivers of the Atlantic coast of Nova Scotia and southwest New Brunswick in Salmon Fishing Areas (SFAs) 19–21 and the portion of SFA 23 outside the inner Bay of Fundy comprise the Scotia-Fundy stocks. With the exception of at least one stock in SFA 19, they have a large salmon component that migrates to the North Atlantic/Labrador Sea (Amiro *et al.*, 2008). Estimates of returns and spawning escapement for the Scotia-Fundy stocks are provided as inputs to the runreconstruction model. Inner Bay of Fundy Atlantic Salmon (SFA 22 and part of SFA 23) have been federally listed as endangered under the Canadian Species at Risk Act and are not included as inputs into the run-reconstruction model. With the exception of one population, inner Bay of Fundy stocks have a localized migration strategy while at sea and an incidence of maturity after one winter at sea.

Consistent with the requirements of the model, a range (minimum to maximum) of returns and spawning escapement for the Scotia-Fundy stocks is provided for the runreconstruction model. The methods used to estimate total returns and spawners are described by Amiro et al. (2008). In brief, for SFAs 19-21, the escapement is based on the count of small and large salmon at the Morgan Falls fish-way on the LaHave River from 1970 to the present year, scaled up to the region using the relationship between this count and the recreational catch data for rivers in SFA 19 to 21 from 1970 to 1997 and a catch rate for the LaHave River from 1970 to 1997. Estimates of the returns also include estimates of landings in the commercial salmon fisheries in SFA 19-21 from 1970 to 1983. The model is fitted using maximum likelihood, and the 90% confidence limits are carried forward as the minimum and maximum values. In SFA 23 from 1970 until 1992, estimates of total 1SW and large wild-origin salmon returns are based on the estimated number of returns destined for tributaries above Mactaquac Dam on the Saint John River; this includes in-river and outer-Fundy commercial landings (1970-1971 and 1981–1983), in-river aboriginal harvests (since 1974), and counts at Mactaquac Dam. These estimates are raised by the proportion of the total accessible productive habitat in SFA 23 that is upstream of Mactaquac Dam (0.4-0.6). Hatchery-origin returns were attributed to above Mactaquac Dam only and no hatchery 1SW and MSW returns were estimated for other rivers within SFA 23 (outer Fundy). Since 1993 the estimates of 1SW and MSW returns to the Nashwaak River have been used to estimate the wild production from tributaries of the Saint John River below Mactaquac Dam. The estimated 1SW and MSW returns to the Nashwaak River (above Counting Fence), is raised by the proportion of the total production area accounted for below Mactaquac (0.21–0.3) and then added to the above Mactaquac totals.

#### USA

Total salmon returns and spawners for USA rivers are based on trap and weir catches and for the small rivers in Maine that do not have fish counting facilities, estimates of spawners were based on redd counts.

## Improvements to NAC input data

Modifications to input variables used in assessments for the NAC area are reported by WGNAS in the year in which they are first implemented.

Over recent years, efforts have been made to reduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes). However, the methods used to derive estimates of unreported catch vary markedly between countries. For example, some countries include only illegally caught fish in the unreported catch, while other countries include estimates of unreported catch by legal gear as well as illegal catches in their estimates.

Descriptions of the national approaches used for evaluating unreported catches have been reported at various WGNAS meetings (e.g. ICES, 1996; 2000; 2002; 2010a). In addition, detailed reports describing national procedures for evaluating illegal and unreported catch, and efforts to minimise this, were submitted by parties to NASCO in 2007 in support of a special theme session on this issue. Full details are available at: <a href="http://www.nasco.int/pdf/2007%20papers/CNL(07)26.pdf">http://www.nasco.int/pdf/2007%20papers/CNL(07)26.pdf</a>

Input data commonly rely on rod catches and the practice of catch-and-release has become increasingly important in recent years to reduce levels of exploitation on stocks. As the practice is increasing, WGNAS have previously recommended (ICES,

2010a) that consideration should be given to incorporating mortality associated with this practice in river-specific, regional and national assessments.

The procedures currently used to incorporate catch-and-release and unreported catches into regional, national and international assessments are summarised at Appendix 1 (from ICES, 2010a).

# Biological and other data requirements

As noted previously, many of the 'conventional' data requirements (e.g. marine survey data and commercial cpue) used in the assessment of other commercially important fish species are inappropriate to salmon. A range of biological, catch and exploitation rates and other data pertinent to appropriate stock assessments are however, collected and made available to WGNAS to help inform assessments and to aid in responding to the various questions posed by NASCO.

Appendix 2 of this Stock Annex provides an overview of current and possible future data requirements for Atlantic salmon assessment/ scientific advice. This was compiled at a recent meeting of WGNAS (ICES, 2013) in relation to monitoring requirements under the European Data Collection Framework (DCF) and following a more detailed review of the data requirements under DCF (ICES, 2012c). This table illustrates the type of information collected/available, but is provided for illustrative purposes only. It should be noted that many Atlantic salmon producing countries fall outside the DCF provisions, which only relate to countries within the European Union. Further, Sovereign states are responsible for the regulation of salmon fisheries within their areas of jurisdiction. Formal ICES catch advice is only required for the distant water salmon fisheries, which take salmon originating in rivers of another party.

# Assessment methods

In managing Atlantic salmon fisheries, NASCO has adopted a fixed escapement strategy (Potter, 2001), in recognition of the importance of the spawning stock to subsequent recruitment. Therefore, in managing the distant water fisheries at Faroes and West Greenland, the spawning requirements of the rivers contributing to these fisheries must be defined. Management advice, expressed as allowable harvest (tonnes), is then predicated on a forecast of salmon abundance prior to the fishery such that the spawning requirements of the contributing stocks can be achieved. The provision of catch advice thus proceeds through a number of steps:

The definition of spawning objectives;

The development of a measure of abundance prior to the fishery; i.e. the prefishery abundance or PFA;

A measure of the spawning stock contributing to the PFA;

A model to forecast the PFA;

The development of a risk analysis framework for the catch advice.

These steps are described in detail in the following sections, subdivided as necessary for the different distant water fisheries and the various stock complexes which contribute to the two fisheries (Greenland and Faroes).

# Definition of spawning objectives

## Management objectives and reference points

Conservation limits (CLs) for North Atlantic salmon have been defined by ICES as the stock level that will achieve long-term average maximum sustainable yield (MSY). NASCO has adopted the following definition of CLs (NASCO, 1998): 'The CL is a limit reference point; having populations fall below these limits should be avoided with high probability.'

Atlantic salmon has characteristics of short-lived fish stocks; mature abundance is sensitive to annual recruitment because there are only a few age groups in the adult spawning stock. Incoming recruitment is often the main component of the fishable stock. For such fish stocks, the ICES MSY approach is aimed at achieving a target escapement (MSY  $B_{\text{escapement}}$ , the amount of biomass left to spawn). No catch should be allowed unless there is a high probabaility that this escapement can be achieved. The escapement level should be set so there is a low risk of future recruitment being impaired, similar to the basis for estimating  $B_{\text{pa}}$  in the precautionary approach. In short-lived stocks, where most of the annual surplus production is from recruitment (not growth), MSY  $B_{\text{escapement}}$  and  $B_{\text{pa}}$  might be expected to be similar

It should be noted that this is equivalent to the ICES precautionary target reference points ( $S_{pa}$ ). Therefore, stocks are regarded by ICES as being at full reproductive capacity only if they are above the precautionary target reference point. This approach parallels the use of precautionary reference points used for the provision of catch advice for other fish stocks in the ICES area.

For the assessment of the status of stocks and advice on management of national components and geographical groupings of the stock complexes, where there are no specific management objectives:

- ICES requires that the lower bound of the 90% confidence interval of the current estimate of spawners is above the CL for the stock to be considered at full reproductive capacity (equivalent to a probability of at least 95% of meeting the CL)
- When the lower bound of the confidence limit is below the CL, but the midpoint is above, then ICES considers the stock to be at risk of suffering reduced reproductive capacity.
- Finally, when the midpoint is below the CL, ICES considers the stock to be suffering reduced reproductive capacity.

Ideally, Atlantic salmon should be assessed and managed on the basis of river-specific stock units, the scale corresponding best to the spawner to recruitment dynamic (Chaput, 2012). In reality, this is not the case for the majority of rivers, although efforts are continuing to develop river-specific CLs and assessment protocols and developments are reported annually to WGNAS (e.g. ICES, 2013).

The risk assessment frameworks applied by WGNAS directly evaluate the risk of meeting or exceeding the stock complex objectives. Managers can choose the risk level which they consider appropriate. ICES considers however that to be consistent with the MSY and the precautionary approach, and given that the CLs are considered to be limit reference points and to be avoided with a high probability, then managers should choose a risk level that results in a low chance of failing to meet the CLs. ICES

recommends that the probability of meeting or exceeding CLs for individual stocks should be greater than 95% (ICES, 2012b).

# Reference points in the NEAC area

River-specific CLs have been derived for salmon stocks in some countries in the NEAC area (France, Ireland, UK (England & Wales), Norway and Sweeden). An interim approach has been developed for estimating national CLs for countries that cannot provide one based upon river-specific estimates. This approach is based on the establishment of pseudo stock–recruitment relationships for national salmon stocks (Potter *et al.*, 2004).

The NEAC-PFA run reconstruction model (see below) provides a means of relating estimates of the numbers of recruits to the numbers of spawners. The numbers of 1SW and MSW spawners are converted into numbers of eggs deposited using the proportion of female fish in each age class and the average number of eggs produced per female. The egg deposition in year 'n' is assumed to contribute to the recruitment in years "n+3" to "n+8" in proportion to the numbers of smolts produced of ages 1 to 6 years respectively. These proportions are then used to estimate the 'lagged egg deposition' contributing to the recruitment of maturing and non-maturing 1SW fish in the appropriate years. The plots of lagged eggs (stock) against the 1SW adults in the sea (recruits) are presented as 'pseudo stock-recruitment' relationships for each homewater country or region that is unable to provide river-specific CLs. In countries where with more than one region, the analysis is carried out for each region separately and the resulting estimates are summed to provide a national figure.

As noted previously, ICES currently define the CL for salmon as the stock size that will result in the maximum sustainable yield (MSY) in the long term. However, it is not straightforward to estimate this point on the stock–recruitment relationships established by the national PFA run-reconstruction models, as the replacement line (i.e. the line on which 'stock' equals 'recruits') is not known for these relationships. This is because the stock is expressed as eggs, while the recruits are expressed as adult salmon. To address this, WGNAS has developed a method for setting biological reference points from the national/regional pseudo stock–recruitment datasets (ICES, 2001). This model assumes that there is a critical spawning stock level below which recruitment decreases linearly towards zero and above which recruitment remains constant. The position of this critical stock level is determined by searching for the stock value that provides the line of best fit for the stock and recruitment data provided by the PFA run-reconstruction model as determined by the residual sum of squares. This point is a proxy for Slim and is therefore defined as the CL for the stock, and is indicated by the inflection point in the hockey-stick relationship (e.g. see example at Figure 3.1.2.1).

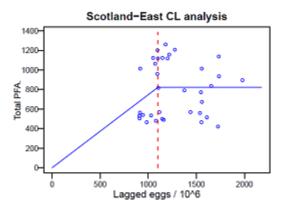


Figure 3.1.2.1. Pseudo stock-recruitment relationship for UK (Scotland) eastern region (from ICES, 2013).

Where river-specific estimates of CLs have been derived for all the rivers in a country or region, these are aggregated to provide national estimates. For countries where the development of river-specific CLs has not been completed, the method described above has been used (see example in Table 3.1.2.1, from ICES, 2013). The estimated national CLs are then summed to provide aggregate CLs for the northern and southern NEAC stock complexes (Table 3.1.2.1).

The CLs have also been used to estimate the spawning escapement reserves (SERs). These represent the CLs increased to take account of natural mortality between the recruitment date, 1st January, and the return to homewaters for maturing and non-maturing 1SW salmon from the northern NEAC and southern NEAC stock complexes (Table 3.1.2.1).

Table 3.1.2.1. Conservation limits (CLs) for NEAC countries and stock complexes estimated from river-specific values, where available, or the national PFA run-reconstruction model. Spawner escapement reserves (SERs) are also included for each stock complex (ICES 2016).

Table 3.2.2.1. Conservation limit options for NEAC stock groups

	National Model CLs		River Spec	River Specific CLs		n limit used	SER		
	1SW	MSW	1SW	MSW	1SW	MSW	1SW	MSW	
Northern Europe		7							
Finland			14,271	9,562	14,271	9,562	17,336	16,386	
Icland (north & east)	5,854	1,678			5,854	1,678	7,218	2,876	
Norway			60,614	72,747	60,614	72,747	77,009	120,991	
Russia	62,752	34,506			62,752	34,506	79,785	61,997	
Sweden			2,099	2,583	2,099	2,583	2,707	4,492	
			Stock Comp	lex	145.590	121.075	184.055	206.742	

-	National Model CLs		River Spe	River Specific CLs		n limit used	SE	SER	
	1SW	MSW	1SW	MSW	1SW	MSW	1SW	MSW	
Southern Europe									
France			17,400	5,100	17,400	5,100	22,440	9,419	
Icland (south & west)	17,790	1,171			17,790	1,171	21,935	2,006	
Ireland			211,471	46,943	211,471	46,943	268,672	78,075	
UK (E & W)			53,988	29,918	53,988	29,918	68,591	51,271	
UK (NI)			19,911	3,280	19,911	3,280	24,365	5,504	
UK (Sco)	256,548	182,741			256,548	182,741	325,942	310,205	
			Stock Comp	lex	577,107	269,153	731,946	456,480	

WGNAS considers the current CL and SER levels may be less appropriate to evaluating the historical status of stocks (e.g. pre-1985), that in many cases have been estimated with less precision.

#### Reference points in the NAC area

In many regions of North America, the CLs are calculated as the number of spawners required to fully seed the wetted area of the river. The methods and values used to derive the egg and spawner conservation requirements for Atlantic Canada are documented in O'Connell et al. (1997). CLs have generally been derived using freshwater production dynamics translated to adult returns to estimate the spawning stock for maximum sustainable yield (MSY). Data were available for a limited number of stocks and these values were transported to the remaining rivers using information on habitat area and the age composition of the spawners. A similar procedure was used to determine the CLs for rivers in the USA (ICES, 1995). In Québec, adult-to-adult stock-recruitment relationships for six rivers were used to define the CLs for the other rivers (Caron et al., 1999). The definition of conservation in Canada varies by region and in some areas, historically the values used were equivalent to maximizing optimizing freshwater production. These are used in Canada as limit reference points and they do not correspond to MSY values. Reference points for Atlantic salmon that conform to the Precautionary Approach have been recently reviewed in eastern Canada (DFO 2015). In 2016, WGNAS provided an overview of ongoing work that Fisheries and Oceans Canada (DFO) was undertaking to refine reference points for Atlantic salmon in Canada that conformed to the Precautionary Approach (ICES 2016a). DFO Maritimes Region (Scotia-Fundy), retained the current conservation requirement based on 240 eggs per 100 m<sup>2</sup> as the Limit Reference Point (DFO 2012; Gibson and Claytor 2013). DFO Newfoundland Region retained the current conservation requirement based on 240 eggs per 100 m² of fluvial rearing habitat, and in addition for insular Newfoundland 368 eggs per ha of lacustrine habitat (or 150 eggs per ha for stocks on the northern peninsula of Newfoundland), as equivalent to their Limit Reference Point and have defined the Upper Stock Reference as 150% of the Limit Reference Point (DFO 2017). The Province of Quebec revised the Limit Reference point and Upper Stock Reference point using a Bayesian hierarchical analysis of stockrecruitment data (Dionne et al., 2015; ICES 2017), and DFO Gulf Region undertook an exercise in 2017 to revise and define the Limit Reference Point in that region of Canada using the proportion of eggs from MSW salmon as a covariate in the Bayesian Hierarchical Model (DFO 2018). The Limit Reference Points in all cases are defined in terms of total eggs from all sizes and sea ages of salmon. Revised reference points specific to 2SW salmon are yet to be developed.

The NAC conservation requirements for 2SW salmon (only these are required in developing catch options for the West Greenland fishery) are summarised in Table 3.1.3.1 (from ICES, 2013). These are calculated from the adult age structure within the different regions and total 123 349 2SW salmon for Canada and 29 199 2SW salmon for the USA, for a combined total of 152 548.

Country and Comission Area	Stock Area	2SW spawner requirement
	Labrador	34 746
	Newfoundland	4022
	Gulf of St Lawrence	30 430
	Québec	29 446
	Scotia-Fundy	24 705
Canada Total		123 349
USA		29 199

152 548

Table 3.1.3.1. 2SW Conservation limits (CLs) for the six regions in the NAC area estimated from river-specific values.

# **Estimating PFA**

North American Total

Estimates of PFA are derived by run-reconstruction methods. These work back in time from estimates of abundance in homewaters to earlier periods of the salmon's life cycle by adding in catches at appropriate times and adjusting for survival. The run-reconstruction approach was first presented at ICES in 1992 and was subsequently adopted for stocks on both sides of the Atlantic (Rago *et al.*, 1993a; Potter and Dunkley, 1993; Potter *et al.*, 1998; 2004). The main advantage of backwards-running, run-reconstruction models over alternative forward-running approaches is that more extensive data are available on adult returns (e.g. traps, counters and catch data) than on freshwater production of juveniles. In addition, rates of natural mortality (M) were thought to be lower and more stable for large salmon after their first winter in the sea than during the post-smolt phase (Potter *et al.*, 2003).

The models used to estimate PFA take the generalised form:

$$PFA = Nh * \exp(Mt_h) + \sum_{i} C_i * \exp(Mt_i)$$

Where: Nh is the number of adult fish returning to homewaters,  $C_i$  the catch of fish from the stock in each interception fishery i (operating before the fish return to homewaters), M the monthly instantaneous rate of natural mortality of salmon in the sea after the first sea-winter,  $t_i$  the time in months between the PFA date and the midpoint of fishery i, and  $t_h$  is the time in months between the PFA date and the midpoint of the return of fish to homewaters. Coastal catches are also added to the estimate where appropriate.

# NEAC area run reconstruction model

The original model used to estimate the PFA of salmon from countries in the NEAC area was described by Potter *et al.* (2004); modifications have been described in subsequent WGNAS reports. PFA in the NEAC area is defined as the number of 1SW recruits on January 1st in the first sea winter. As there are relatively few fish of sea age three or more in most stocks, the model caters for two age groups, 1SW and MSW, the latter including all fish of sea age two or more that are treated as a single cohort. The model is therefore based on the annual catch in numbers of 1SW and MSW salmon in each country. These are raised to take account of minimum and maximum estimates of non-reported catches and exploitation rates of these two sea-age groups.

Thus, for each country (or region) c in year y, the total number of fish of sea age a caught in homewater fisheries ( $Ch_{a,y,c}$ ) is calculated by dividing the declared catch ( $Cd_{a,y,c}$ ) by the non-reporting rate (1 -  $U_{a,y,c}$ ):

$$Ch_{a,y,c} = Cd_{a,y,c} / (1 - U_{a,y,c})$$

where: U<sub>a,y,c</sub> is the estimated proportion of the total catch that is unreported or discarded. The number of fish returning to homewaters (Nh<sub>a,y,c</sub>) is estimated by dividing the total homewater catch by the exploitation rate (H<sub>a,y,c</sub>):

$$Nh_{a,y,c} = Ch_{a,y,c} / H_{a,y,c}$$

As the model provides estimates of total returns and total catch (including non-catch fishing mortality), it is then also possible to estimate the spawner escapement (Nsa,y,e):

$$Ns_{a,y,c} = Nh_{a,y,c} - Ch_{a,y,c}$$

Total catches in the Faroese (Cfa,y) and West Greenland (Cga,y) fisheries are similarly calculated by correcting the declared catches for non-reporting, but they are not raised for the exploitation rate, because the uncaught fish are accounted for from the returns to homewaters. The West Greenland fishery only exploits salmon that would otherwise mature as MSW fish, although the majority are 1SW fish in the summer that they are caught; for the purpose of the model, all are classed as 1SW. The Faroese fishery exploits predominantly MSW salmon, but also a small number of 1SW fish, 78% of which have been estimated to be maturing (ICFS, 1994). Over the past two decades, a substantial proportion of the fish caught in the Faroese fishery have been escapees from salmon farms, and these are discounted from the assessment of wild stocks on the basis of data from Hansen *et al.* (1999). The incidence of farm escapees in the West Greenland catch is thought to be <1.5% (Hansen *et al.*, 1997), so this portion is ignored in the model. The total estimated catches of wild fish in both distant-water fisheries are assigned to the PFA for different countries on the basis of historic tagging studies (Potter, 1996).

The returns to homewaters and catches in the distant-water fisheries of 1SW and MSW salmon are then raised to take account of the marine mortality between January 1st in the first sea winter (the PFA date) and the mid-point of the period over which the respective national fisheries operate. WGNAS determined a natural mortality value of 0.03 per month to be the most appropriate (ICES, 2002) and a range 0.02 to 0.04 is applied within the model in a Monte Carlo simulation. Thus, the PFA of maturing 1SW fish (PFAm), survivors of which will return to homewaters as 1SW adults, is:

```
PFAm_{y,c} = Nh_{1,y,c} * exp(Mt_{h,1,c}) + 0.78 * Cf_{1,y} * w_y * pf_{1,c} * exp(Mt_{f,1,c})
```

and the PFA of non-maturing 1SW fish (PFAn), survivors of which will return to homewaters as MSW adults, is:

```
\begin{split} & PFAn_{y,c} = Nh_{2,y+1,c} * exp(Mt_{h,2,c}) + Cg_{1,y} * pg_{1,c} * exp(Mt_{g,1,c}) \\ & + 0.22 * Cf_{1,y} * w_y * pf_{1,c} * exp(Mt_{f,1,c}) + Cf_{2,y+1} * w_{y+1} * pf_{2,c} * exp(Mt_{f,2,c}) \end{split}
```

where indices y and c represent year and country/region, indices 1 and 2 the 1SW and MSW sea age groups, w is the proportion of the Faroese catch that is of wild origin, pf and pg are the proportion of the catches in the Faroese and West Greenland fisheries originating in each country (as indexed), and th, tf and tg are the times in months between the PFA date and the midpoints of the homewater fisheries, the Faroese

fishery, and the West Greenland fishery, respectively, for the year classes and country/region as indexed.

Total 1SW recruitment for the NEAC area in year y is therefore the sum of the maturing 1SW and non-maturing 1SW recruitments for that year for all countries:

$$PFA_y = \sum_{c} PFAm_{y,c} + \sum_{c} PFAn_{y,c}$$

The non-reporting rates, exploitation rates, natural mortality, and migration times in the above equations cannot be estimated precisely, so national experts provide minimum and maximum values based upon best available knowledge that are considered likely to be centred on the true values (ICES, 2003). A Monte Carlo simulation (12 000 trials) is used to estimate confidence intervals on the stock estimates.

Where appropriate to the provision of management advice, the national outputs from the model are combined into stock complexes, such as those for southern and northern NEAC (ICES, 2002). The confidence limits for these combined estimates are derived from the sum of the national variances obtained from the MCS (the covariances are assumed to be small). This model has provided time-series of PFA estimates for NEAC salmon stocks from 1971 to the present.

The model was initially run using 'Crystal Ball' (CB) in Excel (Decisioneering, 1996). However, an updated version of the model which runs in the 'R' programming language (R Development Core Team, 2007) was developed in 2011 (ICES, 2011). This provided a more flexible platform for the further development of the model and to allow its integration with the Bayesian forecast model for the development of catch options (see below). In 2012, the outputs of the CB and 'R' models were compared to examine the approaches taken and to validate the outputs (ICES, 2012a). Since 2013, the run-reconstruction analysis has been completed by WGNAS using the 'R' programme (ICES, 2013). This has also enabled additional sources of uncertainty to be incorporated into the modelling approach (ICES, 2013).

The full set of country-specific data inputs, as used in the most recent assessment (ICES, 2016) is provided at Appendix 3. The 'R' code used for running the model and the additional data input file required to run the model, are available on the ICES WGNAS SharePoint site.

## NAC area run reconstruction model

The run–reconstruction model developed by Rago *et al.* (1993a) and described in previous WGNAS reports (ICES, 2008; 2009a) and in the primary literature (Chaput *et al.*, 2005) is used to estimate returns and spawners by size (small salmon, large salmon) and sea age group (2SW salmon) to the six geographic regions of NAC from 1971 to the present. The model takes the form:

$$PFA_{vear(i)} = [NR2_{vear(i+1)} * e^{MX1} + NC2_{vear(i+1)}] * e^{MX10} + NC1_{vear(i)} + NG1_{vear(i)}$$

where:  $NR2_{year(i+1)}$  is the sum of 2SW returns to six regions of North America in year i+1,  $NC2_{year(i+1)}$  is the catch of 2SW salmon in Newfoundland and Labrador commercial fisheries in year i+1,  $NC1_{year(i)}$  is the catch of 1SW non-maturing salmon in Newfoundland and Labrador commercial fisheries in year i,  $NG1_{year(i)}$  is the catch of 1SW non-maturing salmon of North American origin in the Greenland fishery in year i, and M is the monthly instantaneous natural mortality of 0.03.

The reconstruction begins with the estimation of returns of 2SW salmon in year i + 1 to six regions in eastern North America: Labrador, Newfoundland, Québec, Gulf, Scotia-Fundy, and USA. For the four southern regions, the regional returns include the harvest in the coastal commercial fisheries but this is not the case for Newfoundland and Labrador. For Labrador, the returns to rivers are estimated from the commercial harvest factored by an exploitation rate. The harvest of 2SW salmon in the Newfoundland and Labrador mixed-stock fisheries in year i + 1 is added to the sum of the returns to the six regions (prorated backward for one month of natural mortality equates to 1st June of year i + 1) to produce the returns to North America. Finally, the harvests of North American origin salmon in the Greenland fisheries in year i and the harvest of non-maturing 1SW salmon in the Newfoundland and Labrador commercial fisheries in year i are added to the prorated returns to North America (ten months between abundance at Greenland on 1st August year i and North America on 1st June year i + 1) to produce the pre-fishery abundance of non-maturing 1SW salmon of North American origin. An instantaneous natural mortality rate of 0.03 per month is assumed for salmon in the second year at sea for all years (ICES, 2002). Adjustments to the input data resulting from reductions and subsequent closures of commercial fisheries in North America are summarized by Friedland et al. (2003).

Following earlier WGNAS recommendations (ICES, 2008), the run–reconstruction model since 2009 has been developed using Monte Carlo simulation (OpenBUGS; <a href="http://mathstat.helsinki.fi/openbugs/">http://mathstat.helsinki.fi/openbugs/</a>; Lunn *et al.*, 2000). This is similar to the approach applied for the NEAC area.

The PFA of the non-maturing component of  $\overline{1}SW$  fish, destined to be 2SW returns (excluding 3SW and previous spawners) is the estimated number of salmon in the North Atlantic on August 1st of the second summer at sea. As this requires estimates of 2SW returns to rivers, there is always a lag in providing this figure (PFA estimates for year n require 2SW returns to rivers in North America in year n + 1).

The full set of data inputs, as used in the most recent assessment (ICES, 2015) is provided at Appendix 4. The 'R' code used for running the model is available on the WGNAS SharePoint site.

# Instantaneous natural mortality rate (M)

The natural mortality rate for salmon after they recruit to the distant water fisheries has been the subject of much discussion. WGNAS originally used a value of 0.01 per month, based upon Doubleday *et al.* (1979), but this was modified to 0.03 per month following a detailed review as part of the EU SALMODEL project (Crozier *et al.*, 2003; ICES, 2002) on the basis of inverse-weight and maturity-schedule models. The rate is assumed to have been constant over the time-series. While mortality may be expected to vary among years and may also be different for maturing and non-maturing 1SW recruits, WGNAS has not had data on which to base the use of different values, or values that change over time. The assumption is, therefore, that the mortality of adult fish after the first winter at sea has not changed and that all the variability of marine mortality has occurred at the post-smolt stage. Efforts are continuing to explore levels of mortality and to better partition this between different stages of the life cycle. The issue was also subject to further investigation within the EU ECOKNOWS project and Bayesian modelling may provide alternative approaches in future.

## PFA forecast models

#### Introduction

The provision of quantitative catch advice for the distant water fisheries requires estimates of abundance before the fisheries take place. While there has been some use of in-season surveys in the management of these fisheries (NASCO, 2001), such methods are considered too impractical and costly to implement on a widespread scale (Potter *et al.*, 2004). Models have therefore been developed by WGNAS which relate abundance estimates obtained at other life stages to the PFA. The objective has been to account for this relationship in terms of biological or environmental factors that affect natural mortality, and to use this to forecast future stock levels.

An initial PFA forecast model for North American stocks (Rago *et al.*, 1993b) utilised indices of thermal habitat in relation to historically observed PFA (from the runreconstruction model) to predict future PFA. Similar approaches were explored by Crozier *et al.* (2003) for NEAC stocks. However, while statistically significant temperature indices could be constructed, the relationships were not always consistent or intuitively correct. Alternative approaches were therefore explored for NEAC; these are described by Potter *et al.* (2004). More recently work by the ICES Study Group on Salmon Stock Assessment and Forecasting (SGSSAFE) has, however, resulted in the development of Bayesian forecast models for both NAC and NEAC (ICES, 2009a; 2011; Chaput, 2012).

In the latest models, PFA dynamics by complex are modelled using the estimates of adult spawners, adjusted to the number of eggs per fish based on life-history characteristics of the age groups within each region of the stock complexes (ICES, 2011; Chaput, 2012). The spawner to PFA dynamic is modelled as:

$$PFA_{y} = e^{\alpha_{y}} LE_{y}e^{\varepsilon}$$

where:  $\alpha_y$  is the productivity parameter from eggs (×1000) to PFA (number of fish) for PFA year y (on a log-scale), LE $_y$  the estimated lagged eggs (×1000) corresponding to the PFA cohort in year y, and the progress of  $\alpha_y$  is modelled as  $a_{y+1} = a_y + \varepsilon$ , with  $\varepsilon \sim N(0, \sigma^2)$ .

Productivity is modelled as an integration of survival in freshwater and during the first year at sea. An important assumption is the absence of heritability of age at maturity, i.e. all eggs are considered equivalent regardless of the age of the spawners. Lagged eggs refer to the adjustment of the egg depositions to correspond to the expected age at smoltification. Spawners in year 'n' contribute to recruitment in years 'n+3' to 'n+8' depending upon the relative proportions of one to six year-old smolts that they produce. For example, spawners in year 'n' produce eggs that hatch in year 'n+1' and may produce one year-old smolts in year 'n+2', which would become 1SW recruits in year 'n+3'. Any two year-old smolts from the same spawners would produce 1SW recruits in year 'n+4', etc.

At the stock complex level, lagged eggs are the sum of the eggs from the spawners in year y - (s + 2) weighted by the proportion of the smolts produced at age s in region k summed over regions in the complex. Two years are added to the smolt age, for the spawning year and smolt migration year, to lag the eggs to the corresponding year of PFA:

$$LE_y = \sum_{k} \sum_{s} Eggs_{y-(s+2),k} * prop_{s,k}$$

#### **NEAC PFA Forecast model**

A forecast model to estimate PFA for all four NEAC stock complexes has been developed in a Bayesian framework by the Study Group on Salmon Stock Assessment and Forecasting (SGSSAFE). The model was originally reported to WGNAS in 2009 (ICES, 2009a), but was subsequently refined and has been in use by WGNAS in its present form since 2011 (ICES, 2011). The models for the northern and southern NEAC stock complexes have exactly the same structure and are run independently. A Directed Acyclical Graph (DAG) for the models is provided in Figure 3.3.2.1. The model considers both the maturing *PFA* (denoted *PFAm*) and the non-maturing *PFA* (denoted *PFAnm*). The full code used for running the model is available on the WGNAS SharePoint site.

A disaggregated version of the Bayesian model has since been developed using the same structure to provide forecasts at a country level, for all countries in both southern and northern NEAC model implementations (ICES, 2013). In these, countries are linked hierarchically only through the variance on the productivity parameter "a". There is no modelling linkage between the northern and southern complexes.

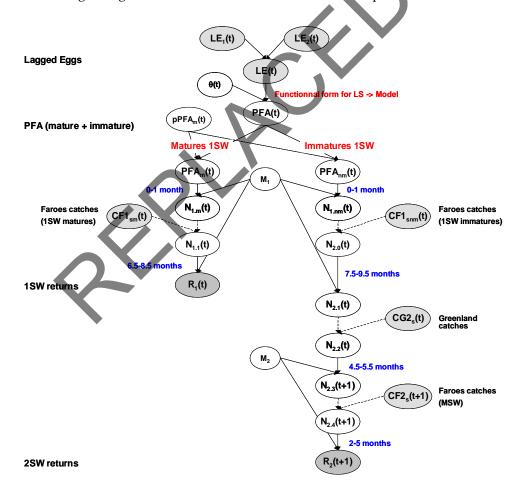


Figure 3.3.2.1. Directed Acyclical Graph (DAG) of the structure of the combined sea age model for the southern NEAC and northern NEAC forecast models. Ellipses in grey are observations (or pseudo-observations) derived from sampling programmes or from submodels (run-reconstruction).

The PFA is modelled using the summation of lagged eggs from 1SW and MSW fish (LE) for each year t and an exponential productivity parameter (a).

```
PFAt = LEt^* exp(at)
```

The productivity parameter (a) is the proportionality coefficient between lagged eggs and PFA. This is forecasted one year at a time ( $a_{t-1}$ ) in an auto correlated random walk, using the previous year's value (a) as the mean value in a normal distribution, with a common variance for the time-series of a.

```
at+1=at+\varepsilon, \varepsilon \sim N(0, a.\sigma 2)
```

The maturing *PFA* (denoted *PFAm*) and the non-maturing *PFA* (denoted *PFAnm*) recruitment streams are subsequently calculated from the proportion of *PFA* maturing (*p.PFAm*) for each year *t. p.PFAm* is forecast as an autocorrelated value from a normal distribution based on a logit scale, using the previous year's value as the mean and a common variance across the time-series of *p.PFAm*.

```
logit.p.PFAmt+1 \sim N(logit.p.PFAmt, p.\sigma2)
logit.p.PFAmt = logit(p.PFAmt)
```

Uncertainties in the lagged eggs were accounted for by assuming that the lagged eggs of 1SW and MSW fish were normally distributed with means and standard deviations derived from the Monte-Carlo run-reconstruction at the scale of the stock complex. The uncertainties in the maturing and non-maturing PFA returns are derived in the Bayesian forecast models through the pseudo-observation method proposed by Michielsens *et al.* (2008), as used in the NAC model.

The natural mortality in the post-PFA time point was assumed constant among years, centred on an instantaneous rate value of 0.03 per month with a 95% confidence interval range of 0.02 to 0.04.

Catches of salmon at sea in the West Greenland fishery (as 1SW non-maturing salmon) and at Faroes (as 1SW maturing and MSW salmon) were introduced as covariates and incorporated directly within the inference and forecast structure of the model. For southern NEAC, the data are available for a time-series of lagged eggs and returns commencing in 1978. Although the return estimates to southern NEAC begin in 1971, the lagged eggs are only available from 1978 due to the smolt age distributions (one to five years). For northern NEAC, data are available for a shorter time-series. Return and spawner estimates begin in 1983, but due to the smolt age distributions (one to six years), the lagged eggs are only available from 1991 onward. The models are fitted and forecasts derived in a consistent Bayesian framework.

The model provides forecasts for maturing and non-maturing stocks for both southern and northern NEAC complexes (and countries) for five years. Risks are defined each year as the posterior probability that the PFA would be greater than or equal to the age and stock complex/ country specific Spawner Escapement Reserves (SERs), under the scenario of no exploitation.

The country disaggregated version of the Bayesian NEAC inference and forecast model incorporates country specific catch proportions at Faroes, lagged eggs and returns of maturing and non-maturing components. Model structure and operation is as described above, incorporating country and year indexing. Linkage between countries in the model is through the common variance parameter associated with the productivity parameter (a) (the proportionality coefficient between lagged eggs and

PFA), which is forecast forward and used along with the forecast proportion maturing to estimate the future maturing and non-maturing PFAs. The evolution of a is independent between countries with the exception of its associated variance. Evolution of the proportion maturing (p.PFAm) is also independent for each country, as is its variance.

#### **NAC PFA Forecast model**

WGNAS (ICES, 2009; 2012; 2015) developed forecasts of the pre-fishery abundance for the non-maturing 1SW salmon (PFA) using a Bayesian framework that incorporates the estimates of lagged spawners and works through the fisheries at sea to determine the corresponding returns of 2SW salmon, conditioned by fisheries removals and natural mortality at sea. This model considers regionally-disaggregated lagged spawners and returns of 2SW salmon for the six regions of North America. The model is summarised in the Directed Acyclical Graph in Figure 3.3.1.1. The year is identified by the i index. The full code used for running the model is available on the WGNAS SharePoint site.

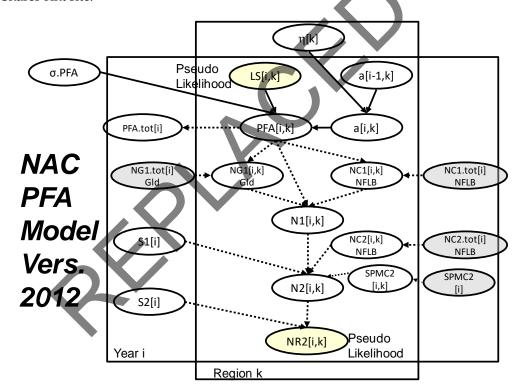


Figure 3.3.1.1. Directed Acyclical Graph (DAG) of the structure of the region disaggregated forecast model for 2SW salmon of North American origin. Ellipses in grey are observations (or pseudo-observations) derived from sampling programmes or from submodels (run-reconstruction).

Lagged spawners LS<sub>i,k</sub> represent the sum of smolt age adjusted annual spawners by region (k) that would be expected to contribute to the recruitment at sea prior to the fisheries (PFA) for year i. LS<sub>i,k</sub> are not directly observed but are estimated from the runreconstruction submodel used to estimates returns and spawners to each of the six regions.

The probability distributions of LS (and returns of 2SW) by region are used as likelihood functions expressing comparative degrees of belief given the data and a probability model not explicitly specified in the current model. The probability distributions were drawn from the Monte Carlo simulations and assumed to be normal

with known mean (LS.m) and precision (1/variance) (tau.LS). The use of this distribution as a likelihood function is equivalent to assuming a pseudo-observation equal to LS.m issuing from a sampling distribution with mean and precision equal to LS and tau.LS (Michielsens *et al.*, 2008).

```
LS.m_{i,k} \sim N (LS_{i,k}, tau.LS_{i,k})
```

The LS.mi,k (mean) and tau.LSi,k (precision) were derived assuming the lagged spawner values issued from a normal distribution characterized by the 95% confidence interval range statistics retained from the Monte Carlo simulations of returns.

Similarly, the returns of 2SW salmon to the six regions (NR2<sub>i,k</sub>) are not directly observed, but estimated from the run-reconstruction model. The probability distributions were assumed to be normal with known mean NR2.m and variance tau.NR2. As with the LS variable, the NR2 were treated as pseudo-observations equal to NR2.m issuing from normal sampling distributions with means and variances equal to NR2 and tau.NR2.

```
NR2.m_{i,k} \sim N (NR2_{i,k}, tau.NR2_{i,k})
```

In between the lagged spawners and returns as 2SW salmon, the catches in the various sea fisheries and conditioning for natural mortality as the fish move from the time of the PFA to homewaters, are incorporated (Figure 3.3.1.1). The catches in the commercial fisheries of West Greenland and the Newfoundland and Labrador commercial and coastal fisheries (NG1.tot. NC1.tot and NC2.tot) are not directly observed, but estimated with error. The catches are converted to numbers of fish of 1SW non-maturing and 2SW fish based on the characteristics of the fish in the catch. Their (prior) probability distributions are obtained from catch statistics according to a formal structure included in the model.

Catches of large salmon (assumed to be 2SW salmon) from the St Pierre & Miquelon fisheries (SPMC) are also included in the model as point estimates.

The natural mortality in the post-PFA time point was assumed constant between years, centred on an instantaneous rate value of 0.03 per month (95% confidence interval range of 0.02 to 0.04).

For the NAC 25W component, the model is fitted to an historical dataseries of lagged eggs starting from 1978. Although the return and spawner estimates for NAC begin in 1971, the lagged eggs are only available from 1978 due to the smolt age distributions (one to six years).

The years are modelled independently conditionally on the lagged spawners and yearly productivity parameters. The lagged spawners to PFA ratios (productivity) are modelled dynamically, i.e. assuming they are sequentially dependent within a region and attempts to take into account the most significant sources of uncertainty. The DAG for the model is shown in Figure 2.

PFA is assumed to be proportional to lagged-spawners (LS), with i.i.d. lognormal errors, and is modelled separately for each region (k = 6). The first year in the timeseries (t) is 1978 for lagged spawners (due to the range of smolt ages 1 to 6 for NAC and the start of the spawner time-series in 1970) and the last year of lagged spawner data is for the 2020 PFA year. The PFA can be modelled for 1978 to 2016 (the last PFA year for which returns of 2SW salmon have been estimated back to rivers in 2017).

$$\begin{aligned} & PFA_{i,k} = LogN(\overline{PFA_{i,k}}, \sigma.PFA^2) \\ & \overline{PFA_{i,k}} = \log(LS_{i,k}) + a_{i,k} + \varepsilon_{i,k} \\ & \varepsilon_{i,k} \quad \stackrel{iid}{\sim} N(0, \sigma.PFA^2) \end{aligned}$$

The total PFA is calculated as the sum of the regional PFA's (k = 6). The proportion of the total PFA in each region is calculated directly as:

$$p.PFA_{i,k} = PFA_{i,k} / PFA.tot_i$$

A non-informative prior is assumed for  $\sigma.PFA^2$  ( $1/\sigma.PFA^2 \sim gamma(0.01, 0.01)$ 

The proportionality coefficient (log)  $a_{i,k}$  between LS<sub>i,k</sub> and PFA<sub>i,k</sub> for each region is modelled dynamically as a random walk with a year and region residual variation ( $\eta_{i,k}$ ) assumed multivariate normal (MVN). The variance covariance matrix ( $\Sigma$ ) allows for correlations among regional productivity values reflecting that the fish share a common marine environment during part of their life cycle and that there are regional specificities in the evolution of the freshwater or the marine coastal environment.

$$a_{i+1,k} = a_{i,k} + \eta_{i,k}$$
$$\eta_{i,k} \sim MVN(0, \Sigma)$$

The common yearly evolution of a is the mean of annual a across regions:

$$a.y_i \leftarrow mean(a_{i,k})$$

This parameterization of the covariance of the proportionality coefficient differs from an earlier version of the model for which the proportionality (log) coefficient  $a_{i,k}$  between LS<sub>i,k</sub> and PFA<sub>i,k</sub> for each region dynamically as a random walk model with the addition of a regionally common annually varying parameter (e.y<sub>i</sub>).

$$a_{i+1,k} = a_{i,k} + e.y_{i+1} + \omega_{i,k} \quad with \quad \omega_{i,k} \stackrel{i.i.d}{\sim} N(0, a.\sigma_k^2)$$

$$e.y_i \stackrel{i.i.d}{\sim} N(0, y.\sigma^2)$$

The correlation matrix of *a* among the regions is calculated from the covariance matrix:

- 1) the precision matrix is inverted to produce the covariance matrix;
- 2) the covariance matrix is transformed to the correlation matrix.

The positive-definite matrix (T, the precision matrix) is inverted:

```
covariance matrix <- solve(T)
correlation matrix <- cov2cor(b)</pre>
```

The dynamic component of the model requires initialization for the first year (i = 1978) and an uninformative prior is assumed:

$$a_{1,k} \stackrel{i.i.d}{\sim} N(0, 100)$$

The models are fitted and forecasts derived in a consistent Bayesian framework under the OpenBUGS 3.0.3 software (<a href="http://mathstat.helsinki.fi/openbugs/">http://mathstat.helsinki.fi/openbugs/</a>; Lunn et al., 2000).

# Summary of NAC and NEAC forecast models

The data inputs and models currently used by WGNAS for forecasting and provision of catch advice differ between the Commission areas; outline details are summarised in the text table below.

	FORECAST MODELS	
	NAC	NEAC
Data inputs		
Time period of data	1978 on	1978 on for southern NEAC 1991 on for northern NEAC
Spatial aggregation	Separately for six regions of North America	By southern and northern stock complexes & NEAC countries
Age components	2SW salmon component only	1SW and MSW age components
Spawners	Lagged spawners by region for 2SW salmon only	Lagged eggs by sea age component for the southern and northern complexes/ country
Returns	Returns by region of 2SW salmon only	Returns of 1SW and MSW age components by stock complex / country
Model structure		
Spatial aggregation	Spawners and returns of 2SW salmon for six regions	Spawners and returns for two sea age components for the southern and northern NEAC complexes / countries
Dynamic function	Random walk dynamic	Random walk dynamic
	Region-specific recruitment rates linked with an annual recruitment rate variable	Sea age specific recruitment rates linked with a probability of maturing variable
Latent variables of interest	PFA 1SW non-maturing Recruitment rate by region and year	PFA 1SW maturing and PFA 1SW non- maturing by stock complex/ country Recruitment rate by sea age component and the probability of maturing variable
Forecast years	Four years	Five years; i.e. the present year -1, the present year, and the next three years
		(y-1 is a forecast, as the MSW stock component is yet to return).

# The development of a risk analysis framework for catch advice

# Introduction

The provision of catch advice in a risk framework involves incorporating the uncertainty in all the factors used to develop the catch options (ICES, 2002). The ranges in the uncertainties of all the factors will result in assessments of differing levels of precision. The analysis of risk involves four steps:

- 1) identifying the sources of uncertainty;
- 2) describing the precision or imprecision of the assessment;
- 3) defining a management strategy; and
- 4) evaluating the probability of an event (either desirable or undesirable) resulting from the fishery action.

The uncertainties have been identified and quantified in the assessment of PFA for salmon stocks in both the NAC and NEAC areas. NASCO's strategy for the management of salmon fisheries is based upon the principle of ensuring that stocks are above CLs (defined in terms of spawner escapement or egg deposition) with a high probability. The undesirable event to be avoided is that the spawning escapement after the fisheries will be below the CLs.

## Catch advice and risk analysis framework for the West Greenland fishery

A risk framework for the provision of catch advice for the West Greenland fishery has been applied since 2003 (ICES, 2003) and has been subject to a number of subsequent updates. The current procedure is outlined below. This involves estimating the uncertainty in meeting defined management objectives at different levels of catch (catch options). The risk framework has been formally accepted by NASCO.

Two stock complexes are of relevance to the management of the West Greenland fishery; non-maturing 1SW fish from North America and non-maturing 1SW fish from southern NEAC. The risk assessments for the two stock complexes are developed in parallel and then combined at the end of the process into a single summary plot or catch options table. The risk analysis proceeds as illustrated in the flowchart in Figure 3.4.2.1).

The primary inputs to the risk analysis for the complex at West Greenland are:

- PFA forecast for the year of the fishery; PFANA and PFANEAC;
- Harvest level being considered (t of salmon);
- Conservation spawning limits or alternate management objectives; and
- The post-fishery returns to each region.

The risk analysis of catch options incorporates the following input parameter uncertainties: (i) the uncertainty of the pre-fishery abundance forecast, (ii) the uncertainty in the biological parameters used to translate catches (weight) into numbers of salmon, and (iii) the uncertainty in attaining the conservation requirements simultaneously in different regions.

The uncertainty in the PFA<sub>NA</sub> and PFA<sub>NEAC</sub> is accounted for in the forecast approach described above. The number of 1SW non-maturing fish of North American and European origin in a given catch (t) is conditioned by the continent of origin of the fish (prop<sub>NA</sub>, prop<sub>E</sub>), by the average weight of the fish in the fishery (Wt<sub>Allages</sub>), and by the proportion 1SW non-maturing fish in the respective continent of origin catches. These parameters define how many fish originating in North America and Europe are expected in the fishery harvests. For a level of fishery under consideration, the weight of the catch is converted to number of fish of each continent of origin using the following equation:

$$C1SW_c = \frac{t \ X \ propC}{ACF * \ (propNA \ X \ Wt1SW_{NA} + propE \ X \ Wt1SW_E)}$$

where: *C1SWc* is the catch (number of fish) of 1SW salmon originating in continent C (either North America or Europe), *t* is the fishery harvest at West Greenland in kg, *propC* is the proportion of the 1SW salmon harvest which originates from continent C, *Wt1SW<sub>NA</sub>* and *Wt1SW<sub>E</sub>* are the average weight (kg) in the fishery of a 1SW salmon of

North American and European origin, respectively, and *ACF* is the age correction factor by weight for salmon in the fishery which are not at age 1SW.

Since these parameters are not known for the forecast years of interest, they are estimated from previous values. Thus, propna (and propneac as 1 - propna) are drawn randomly from observed values of the past five years taking account of uncertainty due to sample sizes. For the other parameters, it is assumed that the parameters for Wtallages and the proportion non-maturing 1SW in the catch by continent of origin could vary uniformly within the values observed in the past five years.

For a level of fishery under consideration, the weight of the catch is converted to fish of each continent of origin and subtracted from one of the simulated forecast values of PFANA and PFANEAC. The fish that escape the Greenland fishery are immediately discounted by the fixed sharing fraction (Fna) historically used in the negotiations of the West Greenland fishery. The sharing fraction chosen is the 40%:60% West Greenland:North America split. The same sharing arrangement has been used for NEAC stocks (ICES, 2003). [Any sharing fraction could be considered and incorporated at this stage of the risk assessment].

After the fishery, fish returning to homewaters are discounted for natural mortality from the time they leave West Greenland to the time they return to rivers. For North America this is a total of eleven months at a rate of M = 0.03 (equates to 28.1% mortality). For southern NEAC stocks this is a total of eight months at a rate of M = 0.03 (equates to 21% mortality). The fish that survive to North American homewaters are then distributed among the regions based on the regional proportions of lagged spawners for the last five years when estimates of spawners were available. The uncertainty in the regional proportions is characterised by drawing at random from a uniform distribution defined by the minimum and maximum regional ranges from the five years and calculating the average proportion for each of the six regions in North America.

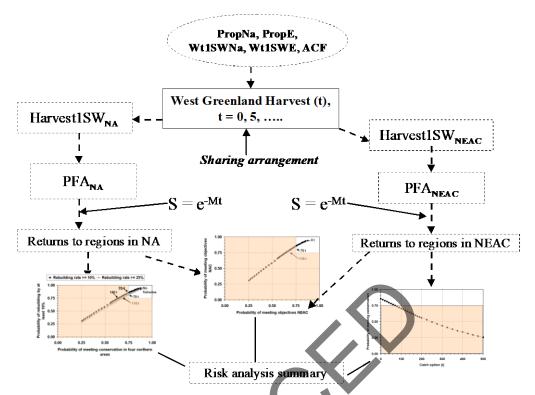


Figure 3.4.2.1. Flow chart summarising risk analysis for catch options at West Greenland using the PFANA and the PFANEAC predictions for the year of the fishery. Inputs with solid borders are considered known without error. Estimated inputs with observation error that is incorporated in the analysis have dashed borders. Solid arrows are functions that introduce or transfer without error whereas dashed arrows transfer errors through the components.

The final step in the risk analysis of the catch options involves combining the conservation requirement or alternate management objectives with the probability distribution of the returns to North America for different catch options. Estimated 2SW returns to each region are compared to the conservation objectives of Labrador, Newfoundland, Québec, and Gulf. Estimated returns for Scotia-Fundy are compared to the objective of achieving an increase of 25% relative to average returns of the base period, 1992–1996. For the USA, the management objective was revised in 2014 (ICES, 2014). Estimated returns for the USA are now compared to the objective of achieving 2SW adult returns of 4549 or greater. The advice to fisheries managers is presented as a probability plot (or table) of meeting or exceeding the objectives relative to increasing harvest levels at West Greenland.

ICES has adopted, a risk level of 75% of simultaneous attainment of seven management objectives (ICES, 2003) as part of an agreed management plan for the West Greenland fishery. The same level of risk aversion is applied for catch advice for homewater fisheries on the North American stock complex.

The catch advice for the West Greenland fishery is currently tabulated to show the probability of each management unit achieving its CL (or alternative reference level) individually and the probability of this being achieved by all management units simultaneously (i.e. in the same given year) (e.g. ICES, 2012a). This allows managers to evaluate both individual and simultaneous attainment levels in making their management decisions. Table 3.4.2.1 provides an example of catch options for West Greenland for the years 2012 to 2014 (ICES, 2012a). An updated catch options table for 2015 to 2017 is provided in Section 5.3 of the 2015 Working Group report (see above).

The models currently used by WGNAS in developing catch advice are considered to provide consistent characterisation of the status and expectations for Atlantic salmon in the North Atlantic. Compared to previous models used by WGNAS prior to 2009, the Bayesian models provide more flexibility, are consistent with the emerging emphasis on such approaches in natural resource assessment, and can provide management advice consistent with the probability of achieving management objectives.

Table 3.4.2.1. Example of catch options tables for mixed-stock fishery at West Greenland by year of PFA, 2015 to 2017.

2015 Catch	Probability of meeting or exceeding region-specific management object								
option (t)	LAB	NFLD	QC	GULF	SF	USA	S-NEAC	ALL	
0	0.45	0.86	0.71	0.50	0.15	0.89	0.98	0.06	
10	0.42	0.84	0.67	0.48	0.14	0.88	0.98	0.05	
20	0.40	0.83	0.63	0.45	0.13	0.87	0.98	0.05	
30	0.38	0.81	0.59	0.42	0.12	0.85	0.98	0.04	
40	0.36	0.78	0.54	0.40	0.12	0.83	0.98	0.04	
50	0.34	0.76	0.50	0.38	0.11	0.81	0.98	0.03	
60	0.32	0.73	0.46	0.36	0.10	0.79	0.98	0.03	
70	0.30	0.70	0.42	0.33	0.09	0.77	0.98	0.03	
80	0.28	0.67	0.39	0.31	0.08	0.74	0.98	0.03	
90	0.26	0.64	0.35	0.29	0.08	0.72	0.97	0.02	
100	0.24	0.60	0.32	0.27	0.07	0.68	0.97	0.02	
2016 Catch	Probabi	lity of meeting	or exceeding	g region-spe	ific manage	ment objecti	ves		
Option (t)	LAB	NFLD	QC	GULF	SF	USA	S-NEAC	ALL	
0	0.48	0.78	0.73	0.50	0.25	0.75	0.95	0.08	
10	0.46	0.76	0.70	0.48	0.24	0.73	0.95	0.07	
20	0.44	0.75	0.67	0.46	0.23	0.72	0.95	0.06	
30	0.42	0.73	0.63	0.44	0.22	0.70	0.95	0.06	
40	0.41	0.70	0.60	0.42	0.21	0.68	0.95	0.06	
50	0.39	0.68	0.56	0.40	0.20	0.66	0.94	0.05	
60	0.37	0.65	0.53	0.38	0.19	0.64	0.94	0.05	
70	0.35	0.63	0.50	0.36	0.18	0.62	0.94	0.05	
80	0.33	0.60	0.47	0.34	0.17	0.59	0.94	0.04	
90	0.31	0.57	0.44	0.32	0.16	0.57	0.94	0.04	
100	0.30	0.54	0.41	0.31	0.15	0.55	0.94	0.04	
20170 Catch	Probabi	lity of meeting	or exceeding	g region-spe	ific manage	ment objecti	ves		
Option (t)	LAB	NFLD	QC	GULF	SF	USA	S-NEAC	ALL	
0	0.56	0.78	0.75	0.55	0.20	0.86	0.94	0.08	
10	0.55	0.77	0.73	0.53	0.20	0.85	0.94	0.08	
20	0.53	0.75	0.70	0.51	0.19	0.84	0.94	0.07	
30	0.52	0.73	0.67	0.49	0.18	0.83	0.94	0.07	
40	0.50	0.71	0.64	0.47	0.17	0.82	0.94	0.06	
50	0.48	0.69	0.62	0.46	0.17	0.81	0.94	0.06	
60	0.46	0.67	0.59	0.44	0.16	0.79	0.94	0.06	
70	0.45	0.65	0.56	0.42	0.16	0.77	0.94	0.05	
80	0.43	0.63	0.54	0.41	0.15	0.76	0.94	0.05	
90	0.42	0.61	0.51	0.39	0.14	0.74	0.94	0.05	
100	0.40	0.59	0.49	0.38	0.14	0.72	0.94	0.05	

## Catch advice and risk analysis framework for the Faroes fishery

#### Outline of the risk framework and management decisions required

There is currently no agreed framework for the provision of catch advice for the Faroes fishery adopted by NASCO. However, NASCO has asked ICES, for a number of years, to provide catch options or alternative management advice with an assessment of risks relative to the objective of exceeding stock conservation limits for salmon in the NEAC area. An initial risk framework that could be used to provide and evaluate catch options for the Faroes fishery was outlined by WGNAS in 2010 (ICES, 2010a). This was based on the method currently used to provide catch advice for the West Greenland fishery, which involves estimating the uncertainty in meeting defined management objectives at different catch levels (TAC options). The Faroes risk framework was developed further at subsequent WGNAS meetings (ICES, 2011; 2012a) and the current proposed procedure is outlined below.

A number of decisions are required by managers before full catch advice could be provided (ICES, 2011; 2012a). Specifically, ICES has indicated that NASCO would need to agree upon the following issues before the risk framework could be finalised:

- season (January to December or October to May) to which any TAC should apply;
- share arrangement for the Faroes fishery;
- choice of management units for NEAC stocks; and
- specification of management objectives.

In developing an indicative risk framework, WGNAS has made pragmatic choices regarding these issues:

Faroes fishing season: A decision is required on the period to which any TAC for the Faroes fishery would apply. The Faroes fishery has historically operated between October/November and May/June, but the historical TACs applied to a calendar year. This means that two different cohorts of salmon of each age class (e.g. two cohorts of 1SW salmon, etc.) were exploited under each TAC. ICES (2011) recommended that NASCO manage any fishery on the basis of fishing seasons operating from October to June, and catch advice should be provided on this basis.

Sharing agreement: The 'sharing agreement' establishes the proportion of any harvestable surplus within the NEAC area that could be made available to the Faroes fishery through the TAC. In effect, for any TAC option being evaluated for the Faroes, it is assumed that the total harvest would be the TAC divided by the Faroes share. WGNAS has proposed using a share allocation derived using the same approach and baseline period (1986–1990) as for West Greenland (ICES, 2010a). This gave a potential share allocation of 7.5% to Faroes. Following discussion within NASCO, one Party proposed an alternative baseline period of 1984–1988, which would give a share allocation of 8.4% to Faroes. In the absence of further advice from NASCO, WGNAS has applied a value of 8.4%.

Choice of management units: ICES (2010a) noted that the stock complexes currently used for the provision of NEAC catch advice (southern NEAC and northern NEAC) are significantly larger than each of the six management units used for North American salmon (2SW only) in the catch advice for the West Greenland fishery. Basing an assessment of stock status on these large units greatly increases the risks to individual

NEAC river stocks or groups of stocks that are already in a more depleted state than the average.

For the provision of catch advice on the West Greenland fishery, the total CL for NAC (2SW salmon only) of about 152 000 fish is assessed in six management units, which means that each unit has an average CL of about 25 000 salmon. In contrast, the total CLs for each of the NEAC stock complexes are:

Northern NEAC 1SW-	158 223
Northern NEAC MSW-	131 356
Southern NEAC 1SW-	565 183
Southern NEAC MSW-	275 549

The NEAC stock complexes are therefore between eight and 25 times the size of the average NAC ones. There is also wide variation in the size and status of stocks both within and among the NEAC national stock groups. WGNAS recommended (ICES, 2012a) that the NEAC catch advice should be based on more management units than are used at present, but also noted that there are practical limitations on the extent to which the assessments can be disaggregated, since the availability of information on the composition of the catch at Faroes constrained the selection of management units. In 2013, WGNAS (ICES, 2013) proposed a method to estimate the stock composition of the Faroes catch at a national level based on tag returns and the PFA estimates, but did not consider it appropriate to extend this to stock complexes smaller than this. Genetic stock assignment studies are underway to analyse scale samples collected at Faroes, but these are not expected to facilitate disaggregation below this level. In addition, other parameter values used in the assessment are currently only available for the total fishery and not smaller stock complexes.

In providing indicative catch advice with the new framework, WGNAS considered that it would be informative to managers to provide catch options tables for the ten NEAC countries as well as for the four stock complexes and has therefore run the risk framework using management units based on countries.

Management objectives: The management objectives provide the basis for determining the risks to stocks in each management unit associated with different catch options. The NASCO agreement on the adoption of a Precautionary Approach (NASCO, 1998) calls for the 'formulation of pre-agreed management actions in the form of procedures to be applied over a range of stock conditions', indicating that the management objectives (e.g. the required probability of exceeding the CL) should be agreed in advance of specific management proposals being considered.

At the request of NASCO, WGNAS considered the implications of applying probabilities of achieving CLs to separate management units vs. the use of simultaneous probabilities; this issue was outlined in detail in ICES (2013).

The probability of simultaneous attainment of management objectives in a number of separate management units is roughly equal to the product of the probabilities of individual attainment for each management unit. The probability threshold for each individual management unit might reasonably be set at a fixed level unless there are specific reasons for adopting an alternative (e.g. for stock rebuilding). ICES (2012) recommended that an appropriate probability level for individual stock complexes would be 95%. This individual probability level can be applied to each management

unit regardless of the number of units used; however, this is less obvious for the probability of simultaneous attainment, as explained next.

Management decisions for the West Greenland fishery have been based on a 75% probability of simultaneous attainment of CLs. For a given probability of achieving individual stock CLs, the probability of simultaneous attainment decreases rapidly as the number of management units considered increases. For the example of 20 management units (e.g. two age groups from each of ten countries), the use of the simultaneous probability level applied for West Greenland (75%) would correspond to the probability of individual stocks meeting the CLs being 98.6% or higher, assuming the same individual probability for all stocks. The use of a 95% probability level for meeting the CLs individually in the 20 management unit example, implies a simultaneous attainment probability of about 36%, i.e. there would be a 64% chance that at least one stock failed to meet its CL in any given year. On the other hand, the use of a 75% probability of simultaneous attainment could result in a fishery being advised when the individual probability of one management unit is as low as 75% if all the other management units have a 100% chance of meeting the CL (as in that case, the probability of simultaneous attainment would still be 75%). This may not be an acceptable risk for managing multiple river stocks.

WGNAS considered that the probability of simultaneous attainment can provide useful information to managers of the risk of failing to meet CLs in at least one stock in the MSF. However, as the management units being considered by NASCO for managing the MSF at Faroes are still very large and each unit encompasses a large number of individual river stocks, choosing a high probability level (such as 95%) of attaining CLs in individual units would be less risky to individual stocks than the use of a simultaneous attainment objective set at the value used for the West Greenland fishery.

On the basis of these considerations, WGNAS provided both individual probabilities and the probability of simultaneous attainment of the management units in the catch options tables (ICES, 2013). ICES recommends that management decisions should be based principally on a 95% probability of attainment of CLs in each stock complex individually. The simultaneous attainment probability may also be used as a guide, but managers should be aware that this will generally be quite low when large numbers of management units are used (as illustrated above, in the example with 20 management units).

# Modelling approach for the catch options risk framework

The basic model for assessing each catch option within the risk framework is the same for both stock complexes and at a country level (ICES, 2013). The PFA forecasts derived in the Winbugs model are transferred to the risk framework model run in 'R'. The estimates and distributions of the PFA estimates used in the risk framework are derived by taking the first 50 000 values from the Winbugs posterior forecast simulations. Parameters in the following description that are marked with an '\*' in the equations have uncertainty around them generated by means of 50 000 random draws from the annual values observed from the sampling programmes conducted in the Faroes between the 1983/1984 and 1990/1991 fishing seasons. They therefore contribute to the estimation of the probability density function around the potential total harvest arising from each TAC option. When the assessment is run at a national level, the number of draws has to be limited to 25 000 because of memory limitations in 'R'.

The modelling procedure involves:

• estimating the total number of 1SW and MSW salmon that could be killed as a result of any TAC at Faroes, including catches in homewaters;

- adjusting these to their equivalent numbers at the time of recruitment to the Faroes fishery;
- subtracting these from the PFA estimates for maturing and non-maturing 1SW salmon in the appropriate years;
- assessing the results against the SERs (i.e. the CLs adjusted to the time of recruitment to the Faroes fishery).

The TAC option (T) is first divided by the mean weight (Wt\*) of salmon caught in the Faroes fishery to give the number of fish that would be caught, and this value is converted to numbers of wild fish (Nw) by multiplying by one minus the proportion of fish-farm escapees in samples taken from the Faroes catch ( $pE^*$ ) observed in historical sampling programmes. A correction factor (C = 0.63) is applied to the proportion of fish-farm escapees to take account of reductions in the numbers of farm escapees over the past 20 years based on observations in Norwegian coastal waters:

$$Nw = T / Wt^* x (1 - (pE^* x C))$$

This value is split into numbers by sea age classes (1SW and MSW) according to the proportion of each age group (pAi\*, where 'i' is 1SW or MSW) observed in historical catch sampling programmes at Faroes. In the past, there has also been a requirement to discard any fish less than 60 cm total length caught in the Faroes fishery, and 80% of these fish were estimated to die, so these mortalities are also added to the 1SW catch. Thus:

```
Nw1SW = Nwtotal \ x \ pA1SW^* + (Nwtotal \ x \ pD^* / (1 - pD^*) \ x \ 0.8) and NwMSW = Nwtotal \ x \ pAMSW^*
```

where: pD\* is the proportion of the total catch that is discarded (i.e. fish <60 cm total length).

Further corrections are made to the 1SW and MSW numbers to reduce the 1SW total to take account of the proportion that will not mature as 1SW fish and to add the survivors from this group to the MSW fish in the following year. For the first catch advice year the number added to the MSW total is adjusted to the TAC of the current season (i.e. zero in 2012/2013). Thus:

```
Nw1SW = Nw1SW \times pK * and NwMSW = NwMSW + Nw1SW \times (1-pK*)
```

where 'pK' is the proportion of 1SW salmon that are expected to mature in the same year (0.78) derived from experimental studies conducted in the 1980s (Youngson and Webb, 1993).

The numbers in each age group are then divided among the management units by multiplying by the appropriate proportions (pUij), where 'i' denotes the age groups and 'j' denotes the management units, and each of these values is raised by the Faroes

share allocation (S) to give the total potential harvest (Hij) of fish from each management unit and sea age group:

$$Nwij = (Nwi \times pUij) / S$$

Finally, these values are adjusted for natural mortality so that they can be compared with the PFA forecasts and SER values from the mid-date of the fishery to the recruitment date by using an instantaneous monthly rate of mortality of 0.03.

These harvests are then subtracted from the stock forecasts (PFAij) for the management units and sea age groups and compared with the Spawner Escapement Reserves (SERij) to evaluate attainment of the management objective. In practice, the attainment of the management objective is assessed by determining the probability that PFAij – Hij – SERij is greater than zero. The SER is the number of fish that need to be alive at the time of the Faroes fishery to meet the CL when the fish return to homewaters; this equals the CL raised by the mortality over the intervening time. CLs and SERs are currently estimated without uncertainty.

## Input data for the risk framework

The analysis estimates probability of each management unit achieving its SER (the overall abundance objective) for different catch options in the Faroes fishery (from 0 to 200 t). The analysis assumes:

- no fishery operated in the 2012/2013 season;
- the TAC allocated to Faroes is the same in each year and is taken in full;
- homewater fisheries also take their catch allocation in full.

The analysis requires the following input data for the catch that would occur at the Faroes if a TAC was allocated (full details are provided in ICES, 2013):

- mean weights;
- proportion by sea age;
- discard rates (fish less than 60 cm total length);
- proportion of fish-farm escapees;
- composition of catches by management unit;
- proportion of 1SW fish not maturing.

# Indicative catch advice

Table 3.4.3.4.1 provides an example of catch options for the Faroes fishery for the seasons 2013/2014 to 2015/2016 (ICES, 2013). Equivalent tables were provided for both 1SW and MSW salmon for all NEAC countries, and WGNAS also estimates the exploitation rates that these TAC options would impose on each stock complex or national stock (ICES, 2013). Updated catch options tables for the seasons 2015/2016 to 2017/2018 are provided in Section 3.5 of the 2015 Working Group report (see above).

Table 3.4.3.4.1. Probability of northern and southern NEAC - 1SW and MSW stock complexes achieving their SERs independently and simultaneously for different catch options for the Faroes fishery in the 2013/2014 to 2015/2016 fishing seasons.

Catch options	TAC option	NEAC-N-	NEAC-N-	NEAC-S-	NEAC-S-	All
for 2013/14	(t)	1SW	MSW	1SW	MSW	complexes
season:	0	96.2%	99.8%	74.3%	75.6%	56.8%
	20	96.2%	99.2%	74.2%	69.8%	52.7%
	40	96.2%	98.2%	74.2%	63.9%	48.2%
	60	96.1%	96.3%	74.1%	57.9%	43.3%
	80	96.1%	93.4%	74.1%	52.1%	38.1%
	100	96.1%	89.3%	74.0%	46.6%	32.9%
	120	96.0%	84.3%	74.0%	41.7%	28.1%
	140	96.0%	78.4%	73.9%	36.8%	23.4%
	160	95.9%	71.6%	73.9%	32.5%	19.2%
	180	95.9%	64.6%	73.8%	28.5%	15.4%
	200	95.8%	57.6%	73.8%	25.0%	12.2%
Catch options	TAC option	NEAC-N-	NEAC-N-	NEAC-S-	NEAC-S-	All
for 2014/15	(t)	1SW	MSW	1SW	MSW	complexes
season:	0	94.6%	99.2%	75.4%	79.6%	59.0%
	20	94.6%	98.2%	75.3%	75.3%	55.8%
	40	94.6%	96.6%	75.3%	70.8%	52.0%
	60	94.5%	94.2%	75.2%	66.4%	48.0%
	80	94.4%	90.9%	75.2%	61.8%	43.6%
	100	94.4%	86.8%	75.1%	57.3%	38.9%
	120	94.3%	82.1%	75.1%	53.1%	34.4%
	140	94.3%	76.8%	75.0%	49.0%	30.1%
	160	94.3%	71.2%	75.0%	45.0%	25.9%
	180	94.2%	65.5%	74.9%	41.5%	22.1%
	200	94.2%	59.6%	74.9%	38.0%	18.6%
Catch options	TAC option	NEAC-N-	NEAC-N-	NEAC-S-	NEAC-S-	All
for 2015/16	(t)	1SW	MSW	1SW	MSW	complexes
season:	0	94.6%	98.5%	70.1%	79.7%	55.2%
	20	94.6%	97.2%	70.1%	76.0%	52.4%
X	40	94.5%	95.1%	70.0%	72.2%	49.2%
	60	94.5%	92.3%	70.0%	68.4%	45.6%
	80	94.5%	89.0%	69.9%	64.6%	41.9%
	100	94.4%	85.0%	69.9%	60.7%	38.0%
	120	94.4%	80.6%	69.8%	57.1%	34.2%
	140	94.3%	75.7%	69.8%	53.5%	30.4%
	160	94.3%	70.6%	69.7%	50.0%	26.7%
	180	94.2%	65.4%	69.7%	46.8%	23.4%
	200	94.2%	60.4%	69.7%	43.7%	20.4%

# Development of indicator frameworks to identify significant changes in previously provided multiannual management advice

## **Background**

In support of the multiannual management advice that is provided for all three NASCO Commission Areas, NASCO asked ICES to provide an assessment of the minimal information needed to signal an unforeseen change in productivity for stocks contributing to fisheries within each Commission area. A particular concern was that an increase in productivity may alter the reliability of the previously provided multiyear catch options and could result in unrealised harvest within various mixed-stock fisheries. Initial progress on this issue was presented to WGNAS in 2006 (ICES, 2006) and further developments were made by the Study Group on Establishing a Framework of Indicators of Salmon Stock Abundance [SGEFISSA] which met in November 2006 (ICES, 2007b) and reported to WGNAS in 2007 (ICES, 2007a). This resulted in the development of a suggested framework (Framework of Indicators - FWI) which could be used to indicate if any significant change in the status of stocks had occurred and thus confirm whether the previously provided multi-annual management advice was still appropriate.

The initial FWI was developed with both the Greenland and Faroes fisheries in mind, although the methodology only proved suitable for the West Greenland fishery and an alternative approach was subsequently developed for the NEAC area (ICES, 2011; 2012a; 2013). Thus, FWIs are now routinely applied in the interim (non-assessment) years of multiyear agreements for both NAC and NEAC to facilitate the management of the West Greenland and Faroes fisheries respectively. Both operate according to the timeline outlined in Figure 3.5.1.1. Outline descriptions of the two different schemes are provided below.

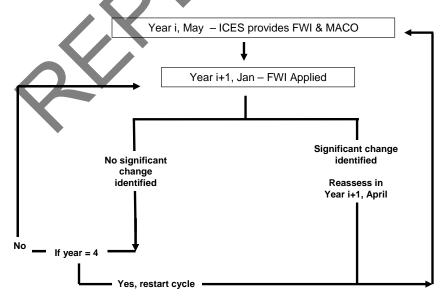


Figure 3.5.1.1. Timeline for employment of the Framework of Indicators (FWI). In Year i, ICES provides an updated FWI which re-evaluates the updated datasets and is summarized in an Excel worksheet. In January of Year i+1 the FWI is applied and two options are available depending on the results. If no significant change is detected, no re-assessment is necessary and the cycle continues to Year i+2. If no significant change is detected in Year i+2, the cycle continues to Year i+3. If a significant change is detected in any year, then reassessment is recommended. In that case, ICES would provide an updated FWI the following May. ICES would also provide an updated FWI if year equals 4. [MACO = multi-annual catch options].

# Framework of Indicators (FWI) for the West Greenland Fishery

The process for developing and applying the FWI for the Greenland fishery consists of six general steps:

- <u>Definition of a significant change</u> Define measurable criteria for what the statement "a significant change in the previously provided multi-annual management advice" represents.
- Evaluating historical relationships between indicators and variables of interest - Define and evaluate the historical relationships between numerous indicators and the variable of interest for individual rivers across all stock complexes.
- <u>Establishing threshold values</u> Define the threshold level (i.e. variable of interest level) that will satisfy the management objectives for each stock complex.
- <u>Decision rule determinations</u> Define and apply a standardised approach for determining the appropriate decision rule value. The decision rule should provide a signal if the variable of interest will be greater than or less than the threshold level with high precision.
- Combining Indicators within the Framework Define and apply a standardised approach for combining indicator datasets within and across stock complexes for future comparison against contemporary indicator values.
- Applying the FWI Define and apply a standard approach to input contemporary indicator values into the FWI to determine if there is likely to be a significant change in the previously provided management advice.

Each of these is considered in turn; full details are available in ICES (2007b).

# Definition of a significant change

A significant change in the previously provided multiannual management advice is regarded as an unforeseen change in stock status that would alter the previously provided advice based on analysis of current population data obtained from various monitored populations across the North Atlantic. This would be indicated by a situation where stock abundance has increased to a level where a fishery could be recommended when no catch had previously been advised, or a decrease in stock abundance when catch options had been chosen.

For the fishery at West Greenland, ICES would recommend that a harvestable surplus exists within the West Greenland stock complex if there was a high probability (75%) that the following three objectives could be met simultaneously:

The conservation limits of the four northern regions of North America (Labrador, Newfoundland, Québec, and Gulf) were achieved.

There was a 25% increase in returns to the Scotia-Fundy region relative to the mean returns for the 1992–1996 period. For the USA, the management objective was revised in 2014 to correspond to recover objectives defined in the recovery plan for endangered Atlantic salmon stocks in the USA (ICES, 2014), this now requires that estimated 2SW adult returns are 4549 or greater.

The conservation limit for the Southern NEAC MSW complex was achieved.

#### Evaluating historical relationships between indicators and variable of interest

A number of variables were considered for inclusion as indicators in the FWI, but only two were considered sufficiently informative to be carried forward into the framework: adult returns (returns, catch or estimated PFA) and return rates (i.e. smolt survival rates, marine survival). These are available, by sea age class, for a number of monitored rivers throughout the North Atlantic and can be directly related to the management objectives for the fishery.

#### Establishing threshold values

In keeping with the 75% probability of meeting or exceeding the objectives for the West Greenland catch options (see above), the 25th percentile of the return estimates of the six areas in North America are compared to the corresponding 2SW conservation limits of the four northern areas of North America, to the 25% increase objective for the Scotia-Fundy area, and to management objective of achieving 4549 or greater 2SW adult returns for the USA. For the southern NEAC non-maturing component, the 25th percentile of the PFA estimate of the southern NEAC non-maturing complex is compared to the spawning escapement reserve (SER) for the southern NEAC non-maturing complex.

#### Decision rule determinations

The procedure for analysing the relationships between the indicators and the returns of 2SW salmon or the PFA estimates was originally suggested by ICES (2006). The individual river catches, returns or return rates are lagged to correspond to the same smolt cohort for the 2SW returns to North America or to the PFA estimates for NEAC complexes. Bivariate plots of each indicator dataset relative to the 2SW returns or the PFA estimates are prepared. Upper and lower halves are defined by the management objective value for the corresponding geographic area in North America or the NEAC stock complexes as outlined above. Estimates of returns of 2SW or PFA estimates in the upper half correspond to years when the returns or PFA exceed the management objectives. Points in the lower half correspond to years when the returns or PFA are less than the management objective.

Left and right halves are defined by a sliding rule along the indicator range. An objective function that maximises the number of correct assignments (true highs and true lows) is used to define the indicator decision rule. The objective function also minimises the number of incorrect assignments (false highs and false lows, Figure 3.5.2.4.1).

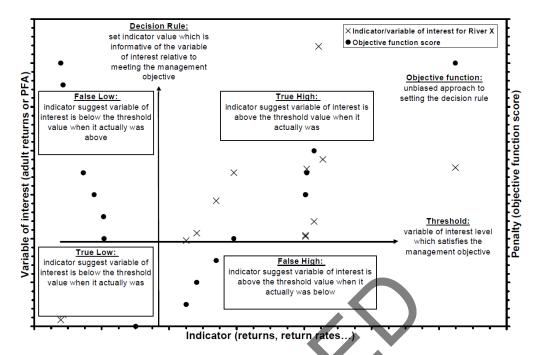


Figure 3.5.2.4.1. Example of Indicator/Variable of Interest exploratory graph identifying the threshold value, decision rule, penalty function and the four states (true high, true low, false high and false low).

The value of the indicator variable that minimises the sum of the penalty scores (i.e. minimises the number of incorrect assignments) is assigned as the decision rule for that dataset. Equal penalty weights are assigned to false highs (lower right quadrant) and false lows (upper left quadrant). Correct assignments are scored as zero. In the case when multiple minima occurred, the lowest indicator value among the low minima values is chosen.

Indicators are retained in the framework when they are evaluated as being informative of the magnitude of returns or PFA relative to the management objectives. These informative indicator datasets also have to meet the following two criteria to be retained:

- Expectation that the indicator variable would be available in future (in January), and
- A minimum of five observations are present in each of the correct quadrats (true low; true high).

# Combining Indicators within the Framework

The probabilities of correct assignments are calculated for each of the true low and true high states for each of the indicator datasets retained. The respective probabilities correspond to the ratio of the correct assignment to all observations within the respective low and high indicator halves:

```
P(State_{low} \mid Indicator_{low}) \ (i.e. \ true \ low) = N(State_{low} \mid Indicator_{low}) \ / \ N Indicator_{low}
```

 $P(State_{high} \mid Indicator_{high}) \ (i.e. \ true \ high) = N(State_{high} \mid Indicator_{high}) \ / \ N$   $Indicator_{high}$ 

Indicator datasets are then pooled according to management objective/stock complex groupings. Each NAC stock complex (n=6) and the southern NEC non-maturing stock complex are pooled separately as these stock complexes relate to the management objectives for the West Greenland fishery.

#### Applying the FWI

To apply the FWI, the most recent year's indicator value for each of the retained indicator datasets is compared to the decision rule as determined from the historical datasets. If the contemporary indicator value is low relative to the decision rule, it is assigned a value of -1. If the value is high, it is assigned a score of +1. Multiple indicators within the stock complex groupings are then combined by arithmetic average of the product of the indicator value (-1, +1) and the probability of a correct assignment corresponding to the true low or true high states. An average geographic area or stock complex score equal to or greater than zero suggests there is a likelihood of meeting the management objective for that grouping based on the historic relation between the variable of interest (adults returns or PFA) and the indicators evaluated.

If the scores for all the groupings within a fishery complex are greater than zero, then there is a likelihood that all the management objectives for that fishery will be met. Under that scenario, the multiyear management advice should be reassessed. When the score(s) for one or more of the groupings is less than zero, there is unlikely to be a significant change in the management advice and there would be no need for a reassessment.

SGEFISSA (ICES, 2007b) developed a spreadsheet template FWI (see example at Figure 3.5.2.6.1) in which the underlying variable of interest/ indicator dataset relationships and decision rules are summarised and collated according to the specific management objectives for each fishery. This provides one of two conclusions for the user:

- 1) No significant change identified by the indicators;
- 2) Reassess

	Catch Advice	Catch opt (Yes = 1,			0	•				
		Overall Recommendation No Significant Change Identified by Indicators								
		2011	Ratio Value to			_	Indicator	Probability of Correct	Indicator	Manageme Objective
	River/ Indicator	Value		Threshold			State	Assignment	Score	Met?
JSA	Penobscot 2SW Returns	2368	167%	1415	100%	92%	1	0.92	0.92	
	Penobscot 1SW Returns Penobscot 2SW Survival (%)	741 0.39	197% 170%	377 0.23	83% 100%	88% 60%	1 1	0.88 0.6	0.88 0.6	
	Penobscot 1SW Survival (%)	0.39	133%	0.23	85%	73%	1	0.6	0.6	
	Narraguagus Returns	196	196%	100	95%	61%	1	0.73	0.73	
	possible range	150	15070	100	-0.93	0.75		0.01	0.01	
	Average		173%						0.75	Yes
Scotia-Fundy	Saint John Return Large	294	9%	3 329	96%	100%	-1	0.96	-0.96	
	Lahave Return Large	146	51%	285	77%	85%	-1	0.77	-0.77	
	St. Mary's Return Large	14	6%	221	100%	73%	-1	1	-1	
	North Return Large	1 193	168%	712	95%		1	0.67	0.67	
	Saint John Return 1SW	582	26%	2 276	86%	80%	-1 -1	0.86	-0.86 -0.94	
	LaHave Return 1SW St. Mary's Return 1SW	565 331	34% 16%	1 679 2 038	94% 95%	67% 93%	-1 -1	0.94 0.95	-0.94 -0.95	
	Saint John Survival 2SW (%)	0.13	59%	0.22	95% 95%		-1 -1	0.95	-0.95 -0.95	
	Lahave Survival 2SW (%)	0.13	367%	0.22	81%	81%	1	0.81	0.81	
	Saint John Survival 1SW (%)	0.12	16%	0.76	86%	73%	-1	0.86	-0.86	
	Lahave Survival 1SW (%)	0.72	50%	1.44	92%	78%	-1	0.92	-0.92	
	Liscomb Survival 2SW (%)	0.03	60%	0.05	86%		-1	0.86	-0.86	
	East Sheet Harbour Survival 2SW (%)	0.005	25%	0.02	67%	82%	-1	0.67	-0.67	
	possible range				-0.88	0.81				
	Average		68%						-0.64	No
	Marriali Barras 2004	00.077	4000/	45.000	4000/	050/		0.05	0.05	
Gulf	Miramichi Return 2SW	28 977 45 880	183%	15 800	100%	85%		0.85	0.85	
	Miramichi Return 1SW possible range	45 660	110%	41 790	89% -0.95	67% 0.76	1 `	0.67	0.67	
	Average		147%		-0.33	0.10			0.76	Yes
						7				
Quebec	Cascapédia Return Large	3 815	167%	2 280	69%	92%	1	0.92	0.92	
	Bonaventure Return Large	1 259	85%	1 479	75%	81%	-1	0.75	-0.75	
	Grande Rivière Return Large	533	121%	442	100%	94%	1	0.94	0.94	
	Saint-Jean Return Large	688	91%	758	86%	89%	-1	0.86	-0.86	
	Dartmouth Return Large	1 171	155%	756	86%	89%	1	0.89	0.89	
	Madeleine Return Large	996	153%	653	70%	93%	1	0.93	0.93	
	Sainte-Anne Return Large	871	201%	433	67%	88%	1	0.88	0.88	
	Godbout Return Large De la Trinite Return Large	694	108% 82%	641	86%	100% 100%	1	1	1 -0.75	
	York Return Return Large	317 1 585	113%	385 1405	75% 63%	83%	-1 1	0.75 0.83	0.83	
	Grande Rivière Return Small	237	119%	199	59%	80%	1	0.83	0.83	
	Saint-Jean Return Small	343	87%	394	53%	80%	-1	0.53	-0.53	
	Godbout Return Small	623	123%	508	85%	92%	1	0.92	0.92	
	De la Trinite Return Small	949	238%	399	89%	83%	1	0.83	0.83	
	De la Trinite Survival Large (%)	0.76	155%	0.49	88%	96%	1	0.96	0.96	
	De la Trinite Survival Small (%)	2.54	170%	1.49	63%	89%	1	0.89	0.89	
	Saint-Jean Survival Small (%)	1.86	258%	0.72	100%	64%	1	0.64	0.64	
	possible range				-0.77	0.88				
	Average		143%						0.50	Yes
	Eurliès Batan Barall	04 005	1070/	04.004	000/	500/		0.50	0.50	
lewfoundland	Exploits Return Small Middle Brook Return Small	34 085 2 642	137% 141%	24 924 1 868	83% 84%	56% 63%	1	0.56 0.63	0.56 0.63	
	Torrent Return Small	2 784	67%	4 154	94%	64%	-1	0.63	-0.94	
	possible range	2 / 04	0776	4 134	-0.87	0.61	-1	0.94	-0.94	
	Average		115%		-0.07	0.01			0.08	Yes
abrador		·								
i uuoi	possible range Average								NA	Unknow
outhern NEAC									NA.	Olikilowi
	possible range								NA	Unknow
	Average								NΑ	UHKNOV

Figure 3.5.2.6.1. Framework of indicators spreadsheet for the West Greenland fishery, updated in April 2018. For illustrative purposes, the 2017 value of returns or survival rates for the 40 retained indicators is entered in the cells corresponding to the annual indicator variable values.

If no significant change has been identified by the indicators, then the multiyear catch advice for the year of interest could be retained. If a significant change is signalled by the indicators, the response is to reassess.

The framework spreadsheet is designed to capture both fishing and non-fishing scenarios:

- Multiyear advice provides no catch options greater than zero but indicators are suggesting that the management objectives may be met (conclusion: Reassess);
- Multiyear advice provides catch options greater than zero but the indicators suggest the management objectives may not be met (conclusion: Reassess).

There are two steps required by the user to run the framework. The first step in the framework evaluation is to enter the catch advice option (i.e. tonnes of catch) for the West Greenland fishery. This feature provides the two way evaluation of whether a change in management advice may be expected and a reassessment would be required. The second step is to enter the values for the indicator variables in the framework for the year of interest. The spreadsheet evaluation update is automated and the conclusion is shown in the row underneath "Overall Recommendation".

The conclusions from the framework evaluation are based on whether there is simultaneous achievement of the management objectives in the six stock areas of North America and the southern NEAC 1SW non-maturing complex. If there are no indicator variables for a geographic area, the attainment of the management objectives is evaluated as unknown and that area or complex is not used in the decision structure of the framework.

Within the geographic areas for which indicator variables are retained, all the available indicators are used to assess the indicator score. If an update value for an indicator variable is not available for the year of interest, the indicator variable is not used to quantify the indicator score for that area.

The West Greenland FWI was updated during the 2018 WGNAS meeting and an updated spreadsheet produced. Details are provided in Section 5.8 of the Working Group report (see above).

## Framework of Indicators (FWI) for the Faroes Fishery

## **Background**

The original FWI applied to the West Greenland fishery (ICES, 2007b) was not applicable for the Faroese fishery for a number of different reasons. Among these were the lack of quantitative catch advice, the absence of specific management objectives and a sharing agreement for this fishery and the fact that none of the available indicator datasets met the criteria for inclusion in the FWI.

In 2011, WGNAS re-evaluated the approach for developing a FWI for the Faroese fishery (ICES 2011). Since over the available time-series the PFA estimates for the NEAC stock complexes have predominately remained above the SER, the Working Group suggested a different set of decision rules for this FWI. It was suggested that the status of stocks should be re-evaluated if the FWI suggests that the PFA estimates are deviating substantially from the median values from the forecast.

# Description of the FWI

It was initially suggested that the 95% confidence interval range for the mean of the indicator prediction, relative to the median forecast value, be used to compute the decision thresholds for whether the indicator suggests a reassessment or not (ICES, 2011). The limits should be computed at the median values of the PFA forecasts in each of the years in multiyear advice. However, the 95% criterion was subsequently reexamined (ICES, 2012a) and it was recommended that the upper and lower 75% confidence limits of the individual predictions be used for comparison (Figure 3.5.3.2.1). WGNAS recognised that this was a relaxation of the decision rule suggested in 2011, and will lead to a larger interval, and thus a lower chance of a reassessment than the approach suggested in 2011. However, this was considered to be a more realistic confidence level given the relatively wide variability of the indicator datasets, and was also consistent with the approach adopted by NAC.

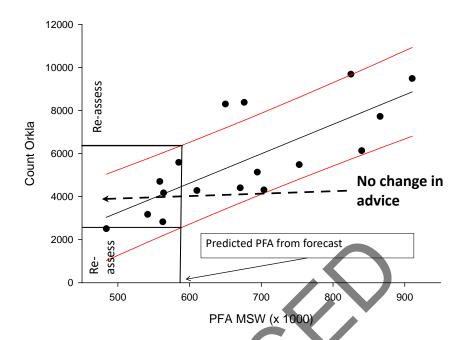


Figure 3.5.3.2.1. Example of how the reassessment intervals for the indicators are computed. The values of an indicator (counts) are plotted against the PFA. The regression line is shown in black and 75% confidence limits for the individual estimates are shown in red. From the forecasted PFA in the year in question the values of the indicator corresponding to the upper and lower 75% confidence interval are estimated. If the indicator value falls outside these limits a reassessment is recommended by this particular indicator.

When the stocks are divided into smaller management units, potential indicators for each management unit become relatively scarce. Therefore, the Working Group recommended that the FWI be regressed against the stock complexes that they belong to. For example, MSW indicators from Norway should be regressed against PFA MSW for northern NEAC.

In 2012, the FWI was applied as a two-tailed test (ICES, 2012a). However, it was subsequently agreed that, in the event of a closed fishery, the indicators should only be compared to the upper 75% confidence limit (i.e. a one-tailed test). This means that for a closed fishery, a reassessment is only triggered where the forecast appeared to be an underestimate and there may be a possibility of a harvest being denied. In the case of an open fishery they should be compared to both the upper and lower 75% confidence limits. In this case, if the FWIs suggest that the forecasted PFA is either an underestimation or an overestimation of the realised PFA in any of the four stock complexes, then this should trigger a reassessment.

ICES further advised (ICES, 2015) that, in the case of closed fisheries, the FWI should only be applied to those stock complexes which had previously signalled zero catch options at Faroes. This was agreed by NASCO (NEA(16)11) and applied in January 2017.

WGNAS developed a FWI spreadsheet (ICES, 2011) to provide an automatic evaluation of the need for a reassessment once the new indicator values are available in January; this has been updated in subsequent years (ICES, 2012a; 2013, 2015). An example spreadsheet is provided at Figure 3.5.3.2.2.

The following summarizes the main steps performed by the spreadsheet following updating of the relevant data for the variable of interest by adding the latest year's number:

- Regression analysis with the dataset *x* to determine its power to predict PFA in the forecasted years.
- Calculation of the 75% confidence intervals of individual predictions of the regression for dataset *x*. An indicator value below the 75% individual confidence interval (CI) is interpreted as indicative of an overestimation of the PFA, while a point above the 75% individual confidence interval is interpreted as indicative of an underestimation of PFA.
- A dataset is considered informative and should be kept as an indicator in the FWI if the following conditions are met: sample size  $(n) \ge 10$ ;  $r^2 \ge 0.2$ ; dataset updated annually and new value available by January 15. Datasets that do not meet these criteria are discarded.
- Apply a binary score to each indicator value. Thus, for dataset *x*, if the current year's indicator value is outside the 75% individual regression point estimate CI (below or above) then that indicator receives a score of 1. If the indicator is within the 75% CI, then the indicator receives a score of -1. In the absence of an indicator datapoint for any year, a score of zero is applied. Whether the indicator value is above or below the upper and lower CI values is checked separately in two spreadsheet columns and a decision whether the indicator value is within the CI is assessed by combining the information in the two columns.
- Separate columns are used to sum the scores for all the indicator datasets within each stock complex. This is done separately for points that fall above the CI and those that fall below. In the case of a two-sided approach (open fishery), if the sum of these columns is ≥0, then the spreadsheet signals "REASSESS"; if the sum is <0, then it signals "No significant OVERestimation of PFA identified by indicators, do not reassess" for indicator values that fall below the CI, and "No significant UNDERestimation of PFA identified by indicators, do not reassess" for indicator values that are above the CI. In case of a one-sided approach (closed fishery), only underestimation will signal a "REASSESS".
- FWI results are generated for each stock complex (northern NEAC maturing and non-maturing, and southern NEAC maturing and non-maturing). A score of ≥0 for any of these stock complexes would signal a reassessment.

WGNAS reassessed the effects of applying stricter criteria than  $r^2 \ge 0.2$  for inclusion of indicators in the FWI. As stricter criteria are used, the number of indicators included reduces rapidly. It was therefore concluded to keep the criterion of  $r^2 \ge 0.2$  in order to obtain a sufficient number of indicators to be able to use the FWI even in the event of one or more indicators being unavailable by the time the FWI is applied each year. The  $r^2$  value of 0.2 corresponds to a value slightly lower than what is considered to be a "large" effect size (r = 0.5,  $r^2 = 0.25$ ) by Cohen (1988). Although a criterion of  $r^2 \ge 0.2$  gives each indicator little predictive power alone (Prairie, 1996), the approach of using a suite of indicators is more similar to meta-analysis (Rosenthal, 1984) meaning that the outcome of the FWI is not dependent on the result of one indicator in isolation, but rather on the combined performance of the indicator set.

The Faroes FWI was updated during the 2015 WGNAS meeting and an updated spreadsheet produced. Details are provided in Section 3.7 of the Working Group report (see above). In evaluating all the time-series, it was noted that the lower 12.5 % CL, which is used to determine which indicator values are outside the 75% CI on the lower side, was negative for some regression relationships for predicted PFA values in 2015 and 2016. Since this would invalidate the use of such indicators (they would not indicate that predicted PFA values are overestimates regardless of how small they are), an additional (fourth) criterion was established as a requirement for including time-series in the FWI. This requires that the lower 12.5% confidence limit for an indicator time-series should be positive for any values of PFA included in the FWI.

The Working Group made further changes to the FWI in 2016 (ICES, 2016). This followed the use of the FWI in January 2016 when the FWI signalled that the PFA of the Northern NEAC MSW stock complex was higher than forecast by the Working Group in 2015 and that a reassessment was necessary. However, in the catch advice provided in 2015 (ICES, 2015), it was the status of the Southern NEAC stock complexes which had resulted in a zero catch option for Faroes. As there was no indication from the FWI analysis that the forecast PFAs for these stocks had been underestimated, a change in the status of the Northern NEAC MSW stock complex alone would not have led to a change in the previous advice. To address this issue, an alternative FWI was developed, where only stock complexes that would be appropriate to changing the multiyear advice were included in the framework in the years between the provision of full catch advice. This revised FWI was adopted by NASCO in 2016.

As future catch advice could be determined by the status of stocks in any of the four stock complexes, it will be necessary to retain indicators for each of these in the FWI. However, in any year, it will only be necessary to apply the indicators from those stock complexes that could result in a change in the multiyear advice following a reassessment.

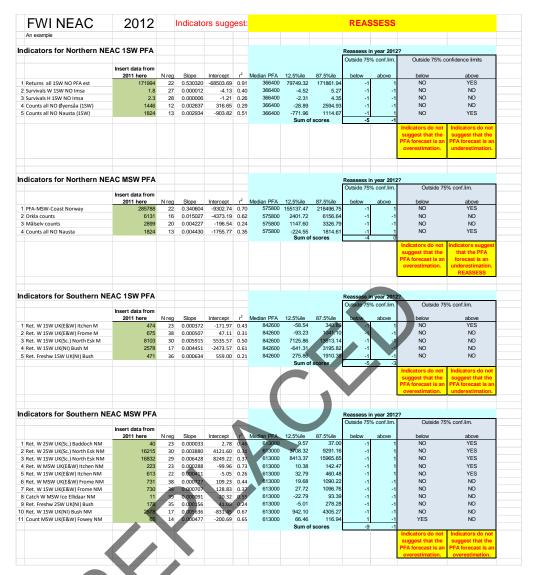


Figure 3.5.3.2.2. Output of the spreadsheet for the test of FWIs for NEAC for 2012 based on the values of the indicators from 2011. Because the indicators suggest that the forecast for northern NEAC MSW was an underestimate, the overall advice from the spreadsheet is reassess.

#### References

Amiro, P., Hubley, B., Gibson, J. and Jones, R. 2008. Estimates of returns and escapements of Atlantic salmon to Salmon Fishing Areas 19, 20 and 21, Nova Scotia, and update of returns and escapements to SFA 23, southern New Brunswick, 1970 to 2007. ICES North Atlantic Salmon Working Group, Working Paper 2008/014.

Anderson, T.C. 1985. The Rivers of Labrador. Can. Spec. Publ. Fish. Aquat. Sci. 81: 389p.

Aprahamian, M.W., Davidson, I.C. and Cove, R.J. 2008. Life history changes in Atlantic salmon from the river Dee, Wales. *Hydrobiologia* 602: 61–78.

Ash, E.G.M. and O'Connell, M.F. 1987. Atlantic salmon fishery in Newfoundland and Labrador, commercial and recreational, 1985. Can. Data Rep. Fish. Aquat. Sci. 672: v + 284 p.

Caron, F., Fontaine, P-M. and Picard, S-E. 1999. Seuil de conservation et cible de gestion pour les rivières à saumon (*Salmon salar*) du Québec. Faune et parcs Québec, Direction de la faune et des habitats. ISBN: 2-550-35636-5, 48 pp.

Cefas and Environment Agency. 2013. Annual assessment of salmon stocks and fisheries in England and Wales, 2012. ISBN: 978 0 907545 74 3, 138pp.

Chaput, G., Legault, C.M., Reddin, D.G., Caron, F. and Amiro, P.G. 2005. Provision of catch advice taking account of non-stationarity in productivity of Atlantic salmon (*Salmo salar L.*) in the Northwest Atlantic. *ICES Journal of Marine Science* 62: 131–143.

- Chaput, G. 2012. Overview of the status of Atlantic salmon (*Salmo salar* L.) in the North Atlantic and trends in marine mortality. *ICES Journal of Marine Science* 69: 1538–1548.
- Cohen, J. 1988. Statistical power analysis for the behavioral sciences. Lawrence Erlbaum Associates, Publishers, Hillsdale, New Jersey.
- Crozier, W.W., Potter, E.C.E., Prevost, E., Schon, P.J., and O'Maoiléidigh, N. 2003. A co-ordinated approach towards development of a scientific basis for management of wild Atlantic salmon in the North-East Atlantic (SALMODEL). An EU Concerted Action Quality of Life and Management of Living Resources Key Action 5: Sustainable agriculture, fisheries and forestry, and integrated development of rural areas including mountain areas. Contract No. QLK5-CT1999-01546, 431 pp.
- Davidson, I.C. and Hazelwood, M.S. 2005. Effects of climate change on salmon fisheries. Science Report W2-047/SR, Environment Agency, Bristol, 52 pp.
- Decisioneering. 1996. Crystal Ball-Forecasting and risk analysis for spreadsheet users (Version 4.0). 286 pp.
- DFO. 2012. Reference points consistent with the precautionary approach for a variety of stocks in the Maritimes Region. DFO Can. Sci. Advis. Sec. Sci. Advis. Rep. 2012/035.
- DFO. 2015. <u>Development of reference points for Atlantic salmon (*Salmo salar*) that conform to the <u>Precautionary Approach</u>. DFO Can. Sci. Advis. Sec. Sci. Advis. Rep. 2015/058.</u>
- DFO. 2017. Stock Assessment of Newfoundland and Labrador Atlantic Salmon 2016. DFO Can. Sci. Advis. Sec. Sci. Advis. Rep. 2017/035.
- DFO. 2018. Limit Reference Points for Atlantic Salmon Rivers in DFO Gulf Region. DFO Can. Sci. Advis. Sec. Sci. Resp. 2018/015.
- Dionne, M., Dauphin, G., Chaput, G., and Prévost, É. 2015. Actualisation du modèle stock-recrutement pour la conservation et la gestion des populations de saumon atlantique du Québec, ministère des Forêts, de la Faune et des Parcs du Québec, Direction générale de la gestion de la faune et des habitats, Direction l'expertise sur la faune aquatique, 66 p.
- Doubleday, W.G., Rivard, D.R., Ritter, J.A., and Vickers, K.U. 1979. Natural mortality rate estimates for North Atlantic salmon in the sea. International Council for the Exploration of the Sea CM 1979/M: 26, 15 pp.
- Fairchild, W.L., Brown, S.B. and Moore, A. 2002. Effects of freshwater contaminants on marine survival in Atlantic salmon. NPAFC: 30–32.
- Friedland, K.D. 1998. Ocean climate influences on critical Atlantic salmon (*Salmo salar*) life history events. *Canadian Journal of Fisheries and Aquatic Sciences* 55 (Suppl. 1): 119–130.
- Friedland, K.D., Reddin, D.G. and Castonguay, M. 2003. Ocean thermal conditions in the post-smolt nursery of North American Atlantic salmon. *ICES Journal of Marine Science* 60: 343–355.
- Friedland, K.D., MacLean, J.C., Hansen, L.P., Peyronnet, A.J., Karlsson, L., Reddin, D.G., Ó Maoiléidigh, N. and McCarthy, J.L. 2009. The recruitment of Atlantic salmon in Europe. *ICES Journal of Marine Science* 66: 289–304.
- Friedland, K.D., Shank, B.V., Todd, C.D., McGinnity, P. and Nye, J.A. 2013. Differential response of continental stock complexes of Atlantic salmon (*Salmo salar*) to the Atlantic Multidecadal Oscillation. Journal of Marine Systems (2013), http://dx.doi.org/10.1016/j.jmarsys.2013.03.003.
- Gibson, A.J.F., and Claytor, R.R. 2013. What is 2.4? Placing Atlantic Salmon Conservation Requirements in the Context of the Precautionary Approach to Fisheries Management in the Maritimes Region. DFO Can. Sci. Advis. Sec. Res. Doc. 2012/043. iv + 21 p.

Glover, K.A., Quintela, M., Wennevik, V., Besnier, F., Sorvik, A.G.E. and Skaala, O. 2012. Three Decades of Farmed Escapees in the Wild: A spatio-temporal analysis of Atlantic salmon population genetic structure throughout Norway. Plos One 7(8): e43129. doi:10.1371/journal.pone.0043129.

- Glover, K.A., Pertoldi, C., Besnier, F., Wennevik, V., Kent, M. and Skaala, O. 2013. Atlantic salmon populations invaded by farmed escapees: quantifying genetic introgression with a Bayesian approach and SNPs. BMC Genetics 14: 74. doi: 10.1186/1471-2156-14-74.
- Hansen, L.P., Døving, K.B. and Jonsson, B. 1987. Migration of farmed adult Atlantic salmon with and without olfactory sense, released on the Norwegian coast. *Journal of Fish Biology* 30: 713–721.
- Hansen, L.P., Jacobsen, J.A. and Lund, R.A. 1997. The incidence of reared Atlantic salmon (*Salmo salar L.*) of fish farm origin at West Greenland. *ICES Journal of Marine Science* 54: 152–155.
- Hansen, L.P. and Quinn, T.P. 1998. The marine phase of Atlantic salmon (*Salmo salar*) life cycle, with comparison to Pacific salmon. *Canadian Journal of Fisheries and Aquatic Sciences* 55 (Suppl. 1): 104–118.
- Hansen, L.P., Jacobsen, J.A. and Lund, R.A. 1999. The incidence of escaped farmed Atlantic salmon, *Salmo salar*, in the Faroese fishery and estimates of catches of wild salmon. *ICES Journal of Marine Science* 56: 200–206.
- Hansen, L.P. and Jacobsen, J.A. 2003. Origin, migration and growth of wild and escaped farmed Atlantic salmon, *Salmo salar* L., in oceanic areas north of the Faroe Islands. *ICES Journal of Marine Science* 60: 110–119.
- Hutchings, J.A. and Jones, M.E.B. 1998. Life history variation and growth rate thresholds for maturity in Atlantic salmon, *Salmo salar*. *Canadian Journal of Fisheries and Aquatic Sciences* 55 (Suppl. 1): 22–47.
- ICES. 1984. Report of the Working Group on North Atlantic Salmon. ICES CM 1984/Assess: 16, 54 pp.
- ICES. 1991. Report of the Working Group on North Atlantic Salmon. Copenhagen, 14–21 March 1991. ICES, CM 1991/Assess: 12, 157 pp.
- ICES. 1993a. Report of the North Atlantic Salmon Working Group. Copenhagen, 5–12 March 1993. ICES, CM 1993/Assess: 10.
- ICES. 1993b. Report of the Study Group on North American Salmon Fisheries. Woods Hole, Mass., 15–19 February 1993. ICES, CM 1993/Assess: 9.
- ICES. 1994. Report of the Working Group on North Atlantic Salmon. Reykjavik, 6–15 April 1994. ICES, Doc. CM 1994/Assess: 16, Ref. M, 182 pp.
- ICES. 1995. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 3–12 April 1995. ICES, Doc. CM 1995/Assess: 14, Ref. M, 191 pp.
- ICES. 1996. Report of the Working Group on North Atlantic salmon, Moncton, Canada, 10–19 April 1996. ICES CM 1996/Assess: 11, Ref: M, 228 pp.
- ICES. 1998. Report of the Working Group on North Atlantic Salmon. Copenhagen, 14–23 April 1998. ICES, CM 1998/ACFM: 15.
- ICES. 1999. Report of the Working Group on North Atlantic Salmon. Copenhagen, 12–22 April 1999. ICES, CM 1999/ACFM: 14.
- ICES. 2000. Report of the Working Group on the North Atlantic Salmon. ICES Headquarters, Copenhagen, April 3–13, ICES CM 2000/ACFM: 13, 301 pp.
- ICES. 2001. Report of the Working Group on North Atlantic Salmon. Aberdeen, 2–11 April 2001. ICES CM 2001/ACFM: 15, 290 pp.
- ICES. 2002. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 3–13 April 2002. ICES CM 2002/ACFM: 14, 299 pp.

ICES. 2003. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 31 March–10 April 2003. ICES CM 2003/ACFM: 19, 297 pp.

- ICES. 2005. Report of the Working Group on North Atlantic Salmon. Nuuk, Greenland 4–14 April. ICES CM 2005/ACFM: 17, 290 pp.
- ICES. 2006. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 4–13 April. ICES CM 2006/ACFM: 23, 254 pp.
- ICES. 2007a. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 11–20 April. ICES CM 2007/ACFM: 13, 253 pp.
- ICES. 2007b. Study Group on Establishing a Framework of Indicators of Salmon Stock Abundance (SGEFISSA). ICES CM 2007/DFC: 01, 71 pp.
- ICES. 2008. Report of the Working Group on North Atlantic Salmon. Galway, Ireland 1–10 April. ICES CM 2008/ACOM: 18, 235 pp.
- ICES. 2009a. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 30 March–8 April 2009. ICES CM 2009/ACFM: 06, 283 pp.
- ICES. 2009b. Report of the Study Group on Biological Characteristics as Predictors of Salmon Abundance (SGBICEPS). ICES CM 2009/DFC: 02, 119 pp.
- ICES. 2010a. Report of the Working Group on North Atlantic Salmon (WGNAS), 22–31 March 2010 Copenhagen, Denmark. ICES CM 2010/ACOM: 09, 302 pp.
- ICES. 2010b. Report of the Study Group on Biological Characteristics as Predictors of Salmon Abundance (SGBICEPS), 24–26 November 2009, ICES Headquarters, Copenhagen, Denmark. ICES CM 2010/SSGEF: 03, 158 pp.
- ICES. 2011. Report of the Working Group on North Atlantic Salmon (WGNAS), 22–31 March 2011 Copenhagen, Denmark. ICES CM 2011/ACOM: 09, 286 pp.
- ICES. 2012a. Report of the Working Group on North Atlantic Salmon (WGNAS), 26 March–4 April 2012, Copenhagen, Denmark. ICES CM 2012/ACOM: 09, 322 pp.
- ICES. 2012b. ICES Advice 2012, Book 10, 99 pp.
- ICES. 2012c. Report of the Workshop on Eel and Salmon DCF Data (WKESDCF). 3–6 July 2012, Copenhagen, Denmark. ICES CM / ACOM: 62, 67 pp.
- ICES. 2013 Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 3–12 April 2013. ICES CM 2013/ACOM: 09. 379 pp.
- ICES. 2014. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 19–28 March 2014. ICES CM 2014/ACOM: 09. 431 pp.
- ICES. 2015. Report of the Working Group on North Atlantic Salmon. Moncton, Canada, 17–26 March 2015. ICES CM 2015/ACOM: 09. 462 pp.
- ICES. 2016. Report of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 30 March-8 April 2016. ICES CM 2016/ACOM: 10. 321 pp.
- IPCC. 2001. IPCC third assessment report: climate change 2001. Synthesis report (ed. R.T. Watson and the Core Writing Team). IPPC, Geneva, Switzerland, 184 pp.
- Jákupsstovu, S.H.í. 1988. Exploitation and migration of salmon in Faroese waters. In Atlantic Salmon: Planning for the Future, pp. 458–482. Ed. by D. Mills and D. Piggins. Croom Helm, London. 587 pp.
- Jacobsen, J.A., Lund, R.A., Hansen, L.P. and O'Maoiléidigh, N. 2001. Seasonal differences in the origin of Atlantic salmon (*Salmo salar* L.) in the Norwegian Sea based on estimates from age structures and tag recaptures. *Fisheries Research* 52: 169–177.
- Jonsson, B. and Jonsson, N. 2009. A review of the likely effects of climate change on anadromous Atlantic salmon *Salmo salar* and brown trout *Salmo trutta*, with particular reference to water temperature and flow. *Journal of Fish Biology* 75: 2381–2447.

Jutila, E., Jokikokko, E. and Julkunen, M. 2006. Long-term changes in the smolt size and age of Atlantic salmon, *Salmo salar* L., in a northern Baltic river related to parr density, growth opportunity and post-smolt survival. *Ecology of Freshwater Fish* 15: 321–330.

- Kennedy, R.J. and Crozier, W.W. 2010. Evidence of changing migratory patterns of wild Atlantic salmon *Salmo salar* smolts in the River Bush, Northern Ireland, and possible associations with climate change. *Journal of Fish Biology* 76, 1786–1805.
- Krkošek, M., Revie, C.W., Gargan, P.G., Skilbrei, O.T., Finstad, B., and Todd, C.D. 2013. Impact of parasites on salmon recruitment in the Northeast Atlantic Ocean. Proceedings of the Royal Society B: Biological Sciences 280: 20122359. http://dx.doi.org/10.1098/rspb.2012.2359
- Lunn, D.J., Thomas, A., Best, N. and Spiegelhalter, D. 2000. WinBUGS-A Bayesian modelling framework: Concepts, structure, and extensibility. Statistics and Computing 10: 325–337.
- MAFF. 1991. Salmon Net Fisheries: Report of a review of salmon net fishing in the areas of the Yorkshire and Northumbria regions of the National Rivers Authority and the salmon fishery districts from the River Tweed to the River Ugie. Pub. HMSO, London.
- Marine Scotland Science. 2010. The North Esk: Scotland's Monitored Salmon River. Topic Sheet 39. URL http://www.scotland.gov.uk/Resource/Doc/295194/0099740.pdf.
- Marine Scotland Science. 2012. Collecting the Marine Scotland salmon and sea trout fishery statistics. Topic Sheet 67. URL http://www.scotland.gov.uk/Resource/0040/00402132.pdf.
- Marine Scotland Science. 2013. Salmon Fishery Statistics. 2012 Season. Topic Sheet 68. URL <a href="http://www.scotland.gov.uk/Resource/0043/00434129.pdf">http://www.scotland.gov.uk/Resource/0043/00434129.pdf</a>.
- Marine Scotland Science. 2017. Assessing conservation status of salmon in Scotland, 28 pp. <a href="http://www.gov.scot/Resource/0052/00524735.pdf">http://www.gov.scot/Resource/0052/00524735.pdf</a>
- Michielsens, C.G.J., McAllister, M.K., Kuikka, S., Mäntyniemi, S., Romakkaniemi, A., Pakarinen, T., Karlsson, L. and Uusitalo, L. 2008. Combining multiple Bayesian data analyses in a sequential framework for quantitative fisheries stock assessment. *Canadian Journal of Fisheries and Aquatic Sciences* 65, 962–974.
- Milner, N. J., Davidson, I.C., Wyatt, R.J., and Aprahamain, M. 2000. The use of spawning targets for salmon fishery management in England and Wales. In: Management and Ecology of River Fisheries, edited by I.G. Cowx, 361-372. Oxford, UK: Blackwell Science.
- Ministère des Forêts, de la Faune et des Parcs. 2016. Plan de gestion du saumon atlantique 2016-2026. Ministère des Forêts, de la Faune et des Parcs, Direction générale de l'expertise sur la faune et ses habitats, Direction de la faune aquatique, Québec, 40 p.
- NASCO. 1994. Explanatory note on the 1994 West Greenland Quota Calculation. NASCO WGC(94), 4 pp.
- NASCO. 1998. North Atlantic Salmon Conservation Organisation. Agreement on the adoption of a precautionary approach. Report of the 15th annual meeting of the Council. CNL(98)46, 4 pp.
- NASCO. 1999. North Atlantic Salmon Conservation Organisation. Action plan for the application of the precautionary approach. CNL(99)48. 14 pp.
- NASCO. 2001. Report of the eighteenth annual meeting of the Commissions, Annex 14. NASCO, Edinburgh. pp. 183–191.
- Nicieza, A.G. and Braña, F. 1993. Relationships among smolt size, marine growth, and sea age at maturity of Atlantic salmon (*Salmo salar*) in northern Spain. *Canadian Journal of Fisheries and Aquatic Sciences* 50: 1632–1640.
- Niemelä, E., Erkinaro, J., Julkunen, M., Hassinen, E., Länsman, M. and Brørs, S. 2006. Temporal variation in abundance, return rate and life histories of previously spawned Atlantic salmon in a large Sub-arctic river. *Journal of Fish Biology* 68: 1222–1240.

O'Connell, M.F., Reddin, D.G., Amiro, P.G., Marshall, T.L., Chaput, G., Mullins, C.C., Locke, A., O'Neil, S.F. and Cairns, D.K. 1997. Estimates of conservation spawner requirements for Atlantic salmon (*Salmo salar* L.) for Canada. Department of Fisheries and Oceans Canadian Stock Assessment Secretariat Research Document 97/100, 58 pp.

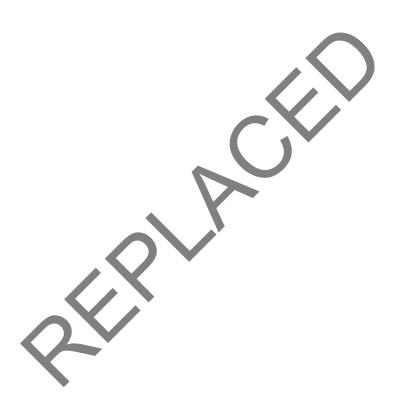
- O'Connell, M.F., Cochrane, N.M., Ash, E.G.M. and Mullins, C.C. 1998. An analysis of the License Stub Return System in the Newfoundland Region, 1994–1997. DFO, CSAS Res. Doc. 98/111, 67 p.
- Otero, J., L'Abée-Lund, J.H., Castro-Santos, T., Leonardsson, K., Storvik, G.O., Jonsson, B., Dempson, J.B., Russell, I.C., Jensen, A.J., Baglinière, J-L., Dionne, M., Armstrong, J.D. Romakkaniemi, A., Letcher, B.H., Kocik, J.F., Erkinaro, J., Poole, R., Rogan, G., Lundqvist, H., Maclean, J.C., Jokikokko, E., Arnekliev, J.V., Kennedy, R.J., Niemelä, E., Caballero, P., Music, P.A., Antonsson, T., Gudjonsson, S., Veselov, A.E., Lamberg, A., Groom, S., Taylor, B.H., Taberner, M., Dillane, M., Arnason, F., Horton, G., Hvidsten, N.A., Jonsson, I.R., Jonsson, N., McKelvey, S., Næsje, T.F., Skaala, O., Smith, G.W., Sægrov, H., Stenseth, N.C. and Vøllestad, L.A. 2014. Basin-scale phenology and effects of climate variability on global timing of initial seaward migration of Atlantic salmon (*Salmo salar*). *Global Change Biology* doi:10.1111/gcb.12363.
- Parrish, D.L., Behnke, R.J., Gephard, S.R., McCormick, S.D. and Reeves, G.H. 1998. Why aren't there more Atlantic salmon (*Salmo salar*)? *Canadian Journal of Fisheries and Aquatic Sciences* 55 (Suppl. 1): 281–287.
- Peyronnet, A., Friedland, K.D., O'Maoiléidigh, N., Manning, M. and Poole, W.R. 2007. Links between patterns of marine growth and survival of Atlantic salmon *Salmo salar*, L. *Journal of Fish Biology* 71: 684–700.
- Peyronnet, A., Friedland, K.D. and O'Maoileidigh, N. 2008. Different ocean and climate factors control the marine survival of wild and hatchery Atlantic salmon *Salmo salar* in the northeast Atlantic Ocean. *Journal of Fish Biology* 73: 945–962.
- Potter, E.C.E. and Dunkley, D.A. 1993. Evaluation of marine exploitation of salmon in Europe. In Salmon in the sea, and new enhancement strategies. Edited by D. Mills. Fishing News Books, Blackwell Science, Oxford. pp. 203–219.
- Potter, E.C.E. 1996. Increases in the returns of salmon to homewaters following the reduction in fishing at Faroes and Greenland. In Enhancement of Spring Salmon. Edited by D. Mills. Proceedings of Atlantic Salmon Trust Conference, London, January 1996. Atlantic Salmon Trust, Pitlochry. pp. 67–86.
- Potter, E.C.E, Hansen, L.P., Gudbergsson, G., Crozier, W.W., Erkinaro, J., Insulander, C., MacLean, J., O'Maoileidigh, N. and Prusov, S. 1998. A method for estimating preliminary conservation limits for salmon stocks in the NASCO-NEAC area. International Council for the Exploration of the Sea CM 1998/T: 17. 11 pp.
- Potter, E.C.E. 2001. Past and present use of reference points for Atlantic salmon. *In*: Stock, Recruitment and Reference Points: Assessment and Management of Atlantic Salmon, pp. 195–223. Ed. by E. Prévost and G. Chaput. INRA Editions, Paris.
- Potter, E.C.E., and Hansen, L.P. 2001. Do Farm Escapees Distort Estimates of Salmon Pre-Fishery Abundance And National Conservation Limits? ICES CM 2001/M:05.
- Potter, E.C.E., Ó Maoiléidigh, N. and Chaput, G. 2003. Marine mortality of Atlantic salmon, *Salmo salar* L: methods and measures. DFO Can. Sci. Adv. Secr. Res. Doc. 2003/101.
- Potter, E.C.E., Crozier, W.W., Schön, P.-J., Nicholson, M.D., Prévost, E., Erkinaro, J., Gudbergsson, G., Karlsson, L., Hansen, L.P., MacLean, J.C., Ó Maoiléidigh, N., and Prusov, S. 2004. Estimating and forecasting pre-fishery abundance of Atlantic salmon (*Salmo salar* L.) in the Northeast Atlantic for the management of mixed stock fisheries. *ICES Journal of Marine Science* 61: 1359–1369.
- Prairie, Y.T. 1996. Evaluating the predictive power of regression models. *Canadian Journal of Fisheries and Aquatic Sciences* 53: 490–492.

R Development Core Team. 2007. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. ISBN 3-900051-07-0, URL http://www.R-project.org.

- Rago, P.J., Meerburg, D.J., Reddin, D.G., Chaput, G.J., Marshal, T.L., Dempson, B., Caron, F., Porter, T.R., Friedland, K.D. and Baum, E.T. 1993a. A continental run reconstruction model for the non-maturing component of North American Atlantic salmon: analysis of fisheries in Greenland and Newfoundland Labrador, 1974–1991. ICES CM 1993/M: 24, 21 pp.
- Rago, P.J., Reddin, D.G., Porter, T.R., Meerburg, D.J., Friedland, K.D. and Potter, E.C.E. 1993b. Estimation and analysis of pre-fishery abundance of the two-sea-winter populations of North American Atlantic salmon (*Salmo salar* L.), 1974–1991. ICES CM 1993/M: 25, 33 pp.
- Reddin, D.G. 2010. Atlantic salmon return and spawner estimates for Labrador. DFO Can. Sci. Advis. Section Res. Doc. 2009/045. iv + 19 p.
- Reddin, D.G. and Dempson, J.B. 1989. Harvest estimates of Sand Hill River Atlantic salmon. Canadian Atlantic Fisheries Scientific Advisory Committee Working Paper 89/104. 19 p.
- Reddin, D.G. and Veinott, G.I. 2010. Atlantic salmon return and spawner estimates for Newfoundland. DFO Can. Sci. Advis. Sec. Res. Doc. 2009/044. iv + 28 p.
- Reddin, D.G., Short, P.B., O'Connell, M.F. and Walsh, A.D. 1996. Atlantic salmon stock status for Sand Hill River, Labrador, 1995. DFO, CSAS Res. Doc. 96/82. 32 p.
- Reddin, D.G., Short, P.B. and Johnson, R.W. 1998. The stock status of Atlantic salmon (*Salmo salar* L.) of Big Brook (Michaels River), Labrador in 1997. DFO, CSAS Res. Doc. 98/119.
- Reddin, D.G., Anthony, R., Watts, K., Nuna, R. and Luther, R.J. 2004. Environmental conditions and harvests in various fisheries for salmonids in Labrador, 2002. CSAS Res. Doc. 2004/003, 20 p.
- Rosenthal, R. 1984. Meta-analytic procedures for social research. Sage Publications, London.
- Royce, W.F. 1984. Introduction to the practice of fishery science. Academic Press, New York.
- Russell, I.C., Aprahamian, M.W., Barry, J., Davidson, I.C., Fiske, P., Ibbotson, A.T., Kennedy, R.J., Maclean, J.C., Moore, A., Otero, J., Potter, E.C.E. and Todd, C.D. 2012. The influence of the freshwater environment and the biological characteristics of Atlantic salmon smolts on their subsequent marine survival. *ICES Journal of Marine Science* 69: 1563–1573.
- Shearer, W.M. 1992. The Atlantic salmon-Natural History, Exploitation and Future Management. Fishing News Books, Oxford, 244 pp.
- Skilbrei, O.T., Finstad, B., Urdal, K., Bakke, G., Kroglund, F. and Strand, R. 2013. Impact of early salmon louse, *Lepeophtheirus salmonis*, infestation and differences in survival and marine growth of sea-ranched Atlantic salmon, *Salmo salar* L., smolts 1997–2009. Journal of Fish Diseases 36: 249-260. doi:10.1111/jfd.12052.
- Smith, G.W., Middlemas, S.J., and MacLean, J.C. 2014. Assessing the status of Scottish Atlantic salmon (Salmo salar L.) stocks using reported catch data: a modelling approach to account for catch and release in the rod and line fishery. Scottish Marine and Freshwater Science 5 (11).
- Todd, C.D., Hughes, S.L., Marshall, C.T., MacLean, J.C., Lonergan, M.E. and Biuw, E.M. 2008. Detrimental effects of recent ocean surface warming on growth condition of Atlantic salmon. *Global Change Biology* 14: 958–970.
- Vøllestad, L.A., Hirst, D., L'Abée-Lund, J.H., Armstrong, J.D., MacLean, J.C., Youngson, A.F. and Stenseth, N.C. 2009. Divergent trends in anadromous salmonid populations in Norwegian and Scottish rivers. *Proceedings of the Royal Society B-Biological Sciences* 276: 1021–1027.
- Vollset, K.W., Barlaup, B.T., Skoglund, H., Normann, E.S. and Skilbrei, O.T. 2014. Salmon lice increase the age of returning Atlantic salmon. Biology Letters 10. DOI: 10.1098/rsbl.2013.0896.

Webb, J.H. 1998. Catch and release: the survival and behaviour of Atlantic salmon angled and returned to the Aberdeenshire Dee, in spring and early summer. Scottish Fisheries Research Report, 62: 1–16.

Youngson, A.F. and Webb, J.H. 1993. Thyroid hormone levels in Atlantic salmon (*Salmo salar*) during the return migration from the ocean to spawn. *Journal of Fish Biology* 42: 293–300.



### **Appendices to Stock Annex**

## Appendix 1 (a): Description of how catch-and-release mortality is incorporated in regional and national stock assessments

Commission			
Area	Country/Region	How it is used in regional and national assessments	Future developments / improvements
NAC		C&R has become more popular in the region and C&R only	New studies of the contribution of C&R fish to spawning
		angling licenses are sold. C&R data are incomplete as there is	success have been initiated. C&R monitoring is becoming
		no requirement to report C&R numbers. Generally, C&R	more complete. Consideration will be given in the future to
		mortality is considered in the assessment but the majority of	incorporating these losses in the returns and in the
		The assessments are conducted as spawner counts after the Fisheries so any losses due to CS Rmortality are accounted for	assessment's based on angling catches, especially as
		in the spawnerestimates but not in the returns (which are the	reporting improves.
		sum of known losses and spawning escapement).	
		and the same and spenting a superiority.	
	Canada-	Catch and release mortality is included in estimates of	No plans for further development.
	Newfoundland	Spawrers. Spawning excapement is reduced by 5-15% (mean	No pains for further development.
	& Labrador	1019 of the released catch.	
	Canada - Gulf	Assessments of spa whers are adjusted by mort ality rates of 3% to 6% of the total CSR estimates of small and large salmon.	Catch and refease mortality is known to be affected by the water temperatures when fish are angled. In some cases,
		The rates vary by river according to angling seasons, and the	angling fisheries are closed when water temperatures are
		occurrence of other factors such as disease which can affect	high in the summer to reduce the losses of fish from C&R.
		survival of salmon.	Methods code termine catch and release numbers vary by
			river and in some same. The marber of released fish is
			estimated from nours and his torical creel survey data. As
			the practice of C&R becomes move opular, estimation
			methods for CS Revolues will have to be revisited.
	Canada –	Assessments are currently adjusted by 4% of the C&R fish to	Numbers of C&R fish are currently low (retention lisheries
	Canada — Scotia/fundy	Assessments are currently adjusted by 4% of the CSR lish to correct for CSR mortality.	closed). If &R catcles is a see, further research on the
			omection at or would be warranted.
	USA	No correction for most ality due to CSR used in estimating	
	USA.	spawrernumbers. However, all fisheries have been dised	
		and the number of fish caught relative tostock size is very	
		small.	
NEAC	Russia	With increasing C&R the retained catch for similar offort is	If the kinformation is incorporated into formal assessments
		reduced. There fore the exploitation rate for retained fish is	the multiple recaptures should be taken into account. C&R
	Norwiy	lower. The increase in C&R in recers years is incorporated into	mortality should be incorporated into estimates of
	Sweden	the national run-reconstruction med. Hey reducing the	spawning e scapement.
	SWIE DE TI	exploitation rate value used in the model input. This is	
	Iceland	assessed qualitative ly. No correction to increased CS.1	
		mort ality is applied when estimating the spowning	
	Ire land	No correction form ortality due to CSR was linest imating	incorporation of formal method for estimating the effect of
		spawner numbers or its the national run-propostruction model.	C&R on number of returning fish, Incorporation of C&R
			mortality in estimates of spawning escapement
	UK(England &	With increasing Call the reserved control or similar effort is	If C&R information is incorporated into formal assessments
	Wales)	reduced. There fore the exploitation rate for retained fish is	then multiple recaptures should be taken into account.
	l .	To be The increase in C&R in resent years is incorporated into the narrowal gars reconstruction model by reducing the	
		exploitation are value used in the model input. This is	
		assessed qualitative ly. 20% mort ality of C&R fish used in	
		as an Ang compliance with river-specific conservation limits.	
	LK(N. Iroland)	Returns are visit imate diby raising the reported net catch by	If C&R information is incorporated into formal assessments
		exploite on rate. No correction for increase dC&R most ality is	then multiple recaptures should be taken into account. C&R
		explired whenestimating the spawning escapement.	mortality should be incorporated into estimates of
		▼	spawring escapement.
	UK(Stotland)	Catch and release mortality is included in estimates of	At rod exploitation rates likely tooccur in UK (Scotland), the
	1	spawners. Spawning escapement is reduced by 10% (± 5%) of	impact of multiple recaptures on abundance estimates is
		the released catch (Webb, 1998; Smith, Middlemas, and	expected to be low (Smith, Middle mas, and Masse an, 2014).
		MarLinan, 2014).	
	Denmark	CRR cate consided but as a storal are assessed to	
	Sermiek	C&R rates recorded, but no national run-reconstruction assessment applied.	
	Finland	No record of C&R	If C&R information is collected, it should be incorporated
	F	N	into formal assessments and multiple recaptures should be
	France	No record of C&R	taken into account. C&R mort ality should be incorporated
NEAC	Evenue		into estimates of spawning escapement.
	Faroes	Notapp	
W. Greenfand	W. Greenland	Notapp	licable

# Appendix 1 (b): Description of how unreported catch is incorporated in regional, national and international stock assessments

Commission						
Area	Country/Region	How it is used in regional and national assessments	How used in international assessments	Future developments / improvements		
NAC	Canada-Quebec	Unreported catches are based on historical estimates relative to stock size or are provided by field conservation and protection staff. Unreported catches when available are included in the regional assessment of returns and spawners.				
	Canada- Newfoundland & Labrador	Catch statistics include estimates of harvests by log book non-respondents. Therefore they are included in the regional assessments and the PFA estimate. No account is taken of illegal fisheries.		If unreported catch estimates were provided they could be incorporated in the regional assessments and in the		
	Canada - Gulf	Unreported catches are sometimes provided by Conservation and Protection Personnel and are estimates of illegal fishing removals within specific regions. Unreported catches have not been used in the assessments of returns or spawners.	Unreported catches which occur in marine waters outside the jurisdiction of the regions are not included in the run reconstruction models.	continent estimates of abundance as spawners. Unreported catch may be accounted for in either the returns of the spawners, depending upon whe and where the illegal activity occurs relative to the location and time of t		
	Canada – Scotia/Fundy	No adjustment made, with the exception of the Saint John River where returns/spawners are adjusted for estimated bycatch and poaching. In other rivers where assessments directly quantify spawners, returns would be underestimated if catch is under reported.		assessment model.		
	USA	Unreported catch is estimated to be zero and therefore has no effect on national assessments.				
NEAC	Russia	Minimum and maximum estimates of the unreporting	National estimates (which incorporate	Incorporate revised estimates of		
	Finland	rate are used in deriving national PFA estimates from	unreported ctaches) are aggregated to	minimum and maximum estimates of		
	Norway	the catch of 1SW & MSW salmon.	provide PFA, return and spawner	unreporting rate as national estimates		
	Sweden		estimates for stock complexes.	are improved.		
	Iceland					
	Ireland					
	UK(England &					
	Wales)					
	UK(N. Ireland)		_			
	France					
	UK(Scotland)	A mean unreporting rate of 10% is applied throughout the time series. Error around this value (s a subjective estimate of the (normally distributed) uncertainty in				
	Daniel.	the parameter and was set to 5%.	_			
NEAC	Denmark Faroes	No national assessment Not applicable	Assumed to be negligible unreported	Sampling programme if fishery		
NEAC	Faroes	Not applicable	catch. Estimate of discard mortality for 1SW fish is incorporated in stock assessments.	resumes.		
W. Greenland	W. Greenland	Not applicable	Unreported catch at West Greenland is incorporated in assessments for both the NAC and NEAC areas. Since 1993,	Annual variation in unreported catch estimates would be incorporated into the model.		
			this has been provided by the Greenlandic authorities. Prior to this time, no unreported catch component is included in the models.			

Appendix 2: Overview of current DCF and future data needs for Atlantic salmon assessment/ scientific advice

			Reviewed and evaluated		Future	
Type of data	Collected under DCF	Available to WG	by WG	Used in current assessment models	plans	Notes
How to be filled	Yes/	Yes/	Yes/	Yes/	Keep as	Free text
	No/	No/	No/	No/	current DCF/	
	Partially	Partially	Partially	Partially used	Improve sampling intensity / No need to be collected / (other free text)	
Fleet capacity	No **	No *	No	No	No need to be collected	See 'Fishing gear and effort'
Fuel consumption	No **	No *	No	No	No need to be collected	Many salmon fisheries use unpowered vessels
Fishing gear and effort	Partially **	Partially	Partially	Partially, but information requested by NASCO	Use for estimation of exploitation rates. Improve coverage and sampling intensity in DC-MAP	Data required for all relevan areas/fisherio s

Type of data	Collected under D	OCF Available to WG	Reviewed and evaluate by WG	ed Used in current assessment models	Future plans	Notes
Landings	Partially **	Yes	Yes	Yes	Improve coverage in DC- MAP	Data required on: catch in numbers and weights for recreational and commercial fisheries in rivers, estuaries and coastal waters.
Discards	No **	No *	No	No	No need to be collected	Not relevant to salmon except (historically) in Faroes fishery. NB: 'catch and release' fish are deliberately caught and so not classed as discards.
Recreational fisheries	Partially **	Yes	Yes	Yes	Improve coverage in DC- MAP	Extent of DCF coverage unclear. Complete catch data needed for all recreational fisheries (see 'Landings')

			Reviewed and evalua		Future		
Type of data	Collected under DCF	Available to WG	by WG	Used in current assessment models	plans	Notes	
Catch & Release	No **	Partially	Partially	No - but data requested by NASCO	Include collectio n in DC- MAP	Data on numbers of fish caught and released required for all recreational fisheries	
cpue dataseries	Partially **	Partially	Partially	Partially	Improve sampling intensity in DC- MAP	Data used to generate national inputs to models	
Age composition	Partially ** Some ageing based on fish lengths or weights	Yes	Yes	Yes	Improve coverage and sampling intensity in DC- MAP	Extent of DCF coverage unclear; sampling intensities in other fisheries inappropriate to salmon	
Wild/reared origin (scale reading)	No **	Partially from other sources	Partially	Partially used - information on farmed fish is requested by NASCO	Improve sampling intensity in DC- MAP	Extent of DCF coverage unclear	
Length and weight-at-age	Partially **	Partially	Yes	Yes - but some ageing based on fish lengths or weights	Improve sampling coverage in DC- MAP	DCF does not cover all relevant areas/fisherie s; sampling intensities inappropriate to salmon	

Type of data	Collected under DCF	Available to WG	Reviewed and evaluated by WG	Used in current assessment models	Future plans	Notes
Sex ratio	No **	Yes- from other sources	Partially	Yes	Modify sampling intensity in DC- MAP	Estimates required at national/regi onal level every five years
Maturity	Not known **	No *	No	No	No need to be collected - all returning adults are mature	DCF requires collection but extent of coverage unclear; data not required for assessments
Fecundity	No **	Yes	Partially	Yes	Include collectio n in DC- MAP	Estimates required at national/regi onal level every five years
Data processing industry	No **	No **	No	No	No need to be collected	Requirement not clear
Juvenile surveys (Electrofishing)	Partially ** but not requested for Atlantic salmon in DCF	Yes	Partially	Partially	Include collectio n in DC- MAP	Data used to develop reference points and confirm stock status. Also required for assessments under WFD

Type of data	Collected under DCF	Available to WG	Reviewed and evaluated by WG	Used in current assessment models	Future plans	Notes
Adult census data (Counters, fish ladders, etc.)	Partially ** but not requested for Atlantic salmon in DCF	Yes	Partially	Yes	Include collectio n in DC- MAP	Counts required for ~one river in 30. Data required to provide exploitation rates for assessments
Index river data (Smolt & adult trapping; tagging programmes; etc.)	Partially ** but not requested for Atlantic salmon in DCF	Yes	Partially	Yes	Include collectio n in DC- MAP	Index rivers are identified by ICES. Data used to develop reference points and inputs to assessment models
Genetic data (for mixed-stock analysis)	No **	Partially	Partially - for some mixed-stock fisheries	Not currently	Include collectio n in DC- MAP - sampling in mixed- stock fisheries every five years	Genetic analysis is now advised to provide more reliable stock composition in mixed- stock fisheries
Economic data	Not known **	No *	No	No - but data are of use to NASCO		Collection of economic data would be useful to managers

Type of data	Collected under DCF	Available to WG	Reviewed and evaluated by WG	Used in current assessment models	Future plans	Notes
Aquaculture data	Not known **	Partially - marine farm production collected	Yes	No - but information on farm production is requested by NASCO		Currently not required for freshwater

Add other data type to the cells with a light blue shading, if needed.

<sup>\*</sup> Not asked for by the ICES WGNAS.

<sup>\*\*)</sup> Not mandatory for some or all areas/stocks/fisheries under the current DCF.

## Appendix 3: Input data for NEAC Pre Fishery Abundance analysis using Monte Carlo simulation

Finland

Annual input data for NEAC PFA run-reconstruction & NCL models for FINLAND (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	8 422	8 538	35.0	10.0	35.0	10.0	50.0	10.0	55.0	15.0
1972	32 789	8 950	35.0	10.0	35.0	10.0	50.0	10.0	55.0	15.0
1973	15 261	14 402	35.0	10.0	35.0	10.0	50.0	10.0	55.0	15.0
1974	21 057	24 508	35.0	10.0	35.0	10.0	50.0	10.0	55.0	
1975	25 242	31 347	35.0	10.0	35.0	10.0	50.0	10.0	55.0	
1976	23 000	24 561	35.0	10.0	35.0	10.0	50.0	10.0	55.0	
1977	12 958	17 035	35.0	10.0	35.0	10.0	50.0	10.0	55.0	
1978	12 338	8 670	35.0	10.0	35.0	10.0	50.0	10.0	55.0	
1979	11 071	7 078	35.0	10.0	35.0	10.0	50.0	10.0	45.0	
1980	10 097	7 994	25.0	10.0	25.0	10.0	50.0	10.0	45.0	15.0
1981	9 049	9 476	25.0	10.0	25.0	10.0	50.0	10.0	45.0	15.0
1982	5 379	12 628	25.0	10.0	25.0	10.0	50.0	10.0	45.0	15.0
1983	13 156	14 013	25.0	10.0	25.0	10.0	50.0	<b>10</b> .0	45.0	
1984	14 371	11 718	25.0	10.0	25.0	10.0	50.0	<b>10</b> .0	45.0	
1985	19 058	11 299	25.0	10.0	25.0	10.0	50.0	10.0	45.0	
1986	15 005	9 320	25.0	10.0	25.0	10.0	50.0	10.0	45.0	
1987	18 151	12 208	25.0	10.0	25.0	10.0	50.0	10.0	45.0	
1988	10 676	8 631	25.0	10.0	25.0	10.0	50.0	10.0	45.0	
1989	27 956	10 337	25.0	10.0	25.0	10.0	60.0	10.0	55.0	
1990	27 955	11 423	25.0	10.0	25.0	10.0	60.0	10.0	55.0	
1991	27 513	15 287	25.0	10.0	25.0	10.0	60.0	10.0	55.0	
1992	38 843	14 826	25.0	10.0	25.0	10.0	60.0	10.0	55.0	
1993	26 195	15 517	25.0	10.0	25.0	10.0	60.0	10.0	55.0	
1994	14 555	14 621	25.0	10.0	25.0	10.0	60.0	10.0	55.0	
1995	14 525	9 625	25.0	10.0	25.0	10.0	60.0	10.0	55.0	
1996	20 466	8 079	25.0	10.0	25.0	10.0	55.0	10.0	50.0	
1997	18 621	9 764	25.0	10.0	25.0	10.0	55.0	10.0	50.0	
1998	23 336	9 307	25.0	10.0	25.0	10.0	55.0	10.0	50.0	15.0
1999	37 495	11 071	25.0	10.0	25.0	10.0	60.0	10.0	50.0	
2000	40 730	21 088	25.0	10.0	25.0	10.0	60.0	10.0	50.0	
2001	29 501	28 112	25.0	10.0	25.0	10.0	60.0	10.0	55.0	
2002	16 721	24 642	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2003	16 497	17 751	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2004	7 002	8 062	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2005	15 366	6 685	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2006	26 916	10 533	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2007	7 862	15 269	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2008	8 481	15 355	25.0	,	25.0	10.0	55.0	10.0	55.0	
2009	15 042	6587	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2010	12 085	10 590	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2011	13 727	8 152	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2012	23 764	9 851	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2013	13 724	9 494	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2014	19 495	10 302	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2015	12 127	9 905	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2016	9 470	10 584	25.0	10.0	25.0	10.0	55.0	10.0	55.0	
2017	4 676	6 645	25.0	10.0	25.0	10.0	50.0	10.0	50.0	10.0

France

Annual input data for NEAC PFA run-reconstruction & NCL models for FRANCE. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Pear   Declared   Catch 15W   Catch MSW   Salmon   Salm	exploitation rate (%) - MSW salmon 12.5 12.5 12.5 12.5 12.5 12.5 12.5
Salmon   Salmon   Salmon   Catch of 1SW   Salmon   Salm	rate (%) - MSW salmon 12.5 12.5 12.5 12.5 12.5 12.5 12.5 12.5
1971   1740   4060   3.5   1.5   37.7	salmon 12.5 12.5 12.5 12.5 12.5 12.5 12.5 12.5
1971       1 740       4 060       3.5       1.5       37.5         1972       3 480       8 120       3.5       1.5       37.5         1973       2 130       4 970       3.5       1.5       37.5         1974       990       2 310       3.5       1.5       37.5         1975       1 980       4 620       3.5       1.5       37.5         1976       1 820       3 380       3.5       1.5       37.5         1977       1 400       2 600       3.5       1.5       37.5         1978       1 435       2 665       3.5       1.5       37.1         1979       1 645       3 055       3.5       1.5       37.1         1980       3 430       6 370       3.5       1.5       37.1         1981       2 720       4 080       3.5       1.5       35.1         1982       1 680       2 520       3.5       1.5       35.1         1983       1 800       2 700       3.5       1.5       35.1         1984       2 960       4 440       3.5       1.5       35.1         1985       1 100       3 330       3.5	12.5 12.5 12.5 12.5 12.5 12.5 12.5
1972       3 480       8 120       3.5       1.5       37.5         1973       2 130       4 970       3.5       1.5       37.5         1974       990       2 310       3.5       1.5       37.5         1975       1 980       4 620       3.5       1.5       37.5         1976       1 820       3 380       3.5       1.5       37.5         1977       1 400       2 600       3.5       1.5       37.5         1978       1 435       2 665       3.5       1.5       37.1         1979       1 645       3 055       3.5       1.5       37.1         1980       3 430       6 370       3.5       1.5       37.1         1981       2 720       4 080       3.5       1.5       35.1         1982       1 680       2 520       3.5       1.5       35.1         1983       1 800       2 700       3.5       1.5       35.1         1984       2 960       4 440       3.5       1.5       35.1         1985       1 100       3 330       3.5       1.5       35.1         1986       3 400       3 400       7.0	12.5 12.5 12.5 12.5 12.5 12.5
1973         2 130         4 970         3.5         1.5         37.5           1974         990         2 310         3.5         1.5         37.5           1975         1 980         4 620         3.5         1.5         37.5           1976         1 820         3 380         3.5         1.5         37.5           1977         1 400         2 600         3.5         1.5         37.5           1978         1 435         2 665         3.5         1.5         37.5           1979         1 645         3 055         1.5         37.5           1980         3 430         6 370         3.5         1.5         37.5           1981         2 720         4 080         3.5         1.5         35.1           1982         1 680         2 520         3.5         1.5         35.1           1983         1 800         2 700         3.5         1.5         35.1           1984         2 960         4 440         3.5         1.5         35.1           1985         1 100         3 330         3.5         1.5         35.1           1986         3 400         3 400         7.0	12.5 12.5 12.5 12.5 12.5
1974         990         2 310         3.5         1.5         37.5           1975         1 980         4 620         3.5         1.5         37.5           1976         1 820         3 380         3.5         1.5         37.5           1977         1 400         2 600         3.5         1.5         37.5           1978         1 435         2 665         3.5         1.5         37.5           1980         3 430         6 370         3.5         1.5         37.5           1981         2 720         4 080         3.5         1.5         37.5           1982         1 680         2 520         3.5         1.5         35.6           1983         1 800         2 700         3.5         1.5         35.6           1984         2 960         4 440         3.5         1.5         35.6           1986         3 400         3 400         3.5         1.5         35.6           1987         6 013         1 806         7.0         5.0         35.6           1988         2 063         4 964         7.0         5.0         35.6           1990         1 886         2 332	12.5 12.5 12.5
1975         1980         4620         3.5         1.5         37.5           1976         1820         3380         3.5         1.5         37.5           1977         1400         2600         3.5         1.5         37.5           1978         1435         2665         3.5         1.5         37.5           1979         1645         3055         3.5         1.5         37.5           1980         3430         6370         3.5         1.5         37.5           1981         2720         4080         3.5         1.5         35.6           1982         1680         2520         3.5         1.5         35.6           1983         1800         2700         3.5         1.5         35.6           1984         2960         4440         3.5         1.5         35.6           1985         1100         3330         3.5         1.5         35.6           1987         6013         1806         7.0         5.0         35.6           1988         2063         4964         7.0         5.0         35.6           1999         1886         2332         7.0 <td< td=""><td>12.5 12.5</td></td<>	12.5 12.5
1976       1 820       3 380       3.5       1.5       37.5         1977       1 400       2 600       3.5       1.5       37.5         1978       1 435       2 665       3.5       1.5       37.5         1979       1 645       3 055       3.5       1.5       37.5         1980       3 430       6 370       3.5       1.5       37.5         1981       2 720       4 080       3.5       1.5       35.1         1982       1 680       2 520       3.5       1.5       35.1         1983       1 800       2 700       3.5       1.5       35.1         1984       2 960       4 440       3.5       1.5       35.1         1985       1 100       3 330       3.5       1.5       35.1         1986       3 400       3 400       7.0       5.0       35.1         1987       6 013       1 806       7.0       5.0       35.1         1988       2 063       4 964       7.0       5.0       35.1         1990       1 886       2 332       7.0       5.0       35.1         1991       1 362       2 125       7.0 <td>12.5</td>	12.5
1977       1 400       2 600       3.5       1.5       37.5         1978       1 435       2 665       3.5       1.5       37.5         1979       1 645       3 055       3.5       1.5       37.1         1980       3 430       6 370       3.5       1.5       37.1         1981       2 720       4 080       3.5       1.5       35.1         1982       1 680       2 520       3.5       1.5       35.1         1983       1 800       2 700       3.5       1.5       35.1         1984       2 960       4 440       3.5       1.5       35.1         1985       1 100       3 330       3.5       1.5       35.1         1986       3 400       3 400       7.0       5.0       35.6         1987       6 013       1 806       7.0       5.0       35.6         1988       2 063       4 964       7.0       5.0       35.6         1989       1 124       2 282       7.0       5.0       35.0         1991       1 362       2 125       7.0       5.0       35.0         1992       2 490       2 671       7.0 <td></td>	
1978       1 435       2 665       3.5       1.5       37.5         1979       1 645       3 055       3.5       1.5       37.5         1980       3 430       6 370       3.5       1.5       37.5         1981       2 720       4 080       3.5       1.5       35.6         1982       1 680       2 520       3.5       1.5       35.6         1983       1 800       2 700       3.5       1.5       35.6         1984       2 960       4 440       3.5       1.5       35.6         1986       3 400       3 400       3.5       1.5       35.6         1987       6 013       1 806       7.0       5.0       35.6         1988       2 063       4 964       7.0       5.0       35.6         1990       1 886       2 332       7.0       5.0       35.6         1991       1 362       2 125       7.0       5.0       35.6         1992       2 490       2 671       7.0       5.0       35.6         1994       2 810       2 290       7.0       5.0       30.0	12.5
1979       1 645       3 055       3.5       1.5       37.5         1980       3 430       6 370       3.5       1.5       37.5         1981       2 720       4 080       3.5       1.5       35.6         1982       1 680       2 520       3.5       1.5       35.6         1983       1 800       2 700       3.5       1.5       35.6         1984       2 960       4 440       3.5       1.5       35.6         1985       1 100       3 330       3.5       1.5       35.6         1986       3 400       3 400       7.0       5.0       35.6         1987       6 013       1 806       7.0       5.0       35.6         1988       2 063       4 964       7.0       5.0       35.6         1990       1 186       2 332       7.0       5.0       35.6         1991       1 362       2 125       7.0       5.0       35.6         1992       2 490       2 671       7.0       5.0       35.6         1994       2 810       2 290       7.0       5.0       30.0	
1980       3 430       6 370       3.5       1.5       37.5         1981       2 720       4 080       3.5       1.5       35.1         1982       1 680       2 520       3.5       1.5       35.1         1983       1 800       2 700       3.5       1.5       35.1         1984       2 960       4 440       3.5       1.5       35.1         1985       1 100       3 330       3.5       1.5       35.1         1986       3 400       3 400       7.0       5.0       35.1         1987       6 013       1 806       7.0       5.0       35.1         1988       2 063       4 964       7.0       5.0       35.1         1989       1 124       2 282       7.0       5.0       35.1         1990       1 886       2 332       7.0       5.0       35.1         1991       1 362       2 125       7.0       5.0       35.1         1992       2 490       2 671       7.0       5.0       35.1         1993       3 581       1 254       7.0       5.0       35.0         1994       2 810       2 290       7.0 <td></td>	
1981       2 720       4 080       3.5       1.5       35.0         1982       1 680       2 520       3.5       1.5       35.0         1983       1 800       2 700       3.5       1.5       35.0         1984       2 960       4 440       3.5       1.5       35.0         1985       1 100       3 330       3.5       1.5       35.0         1986       3 400       3 400       7.0       5.0       35.0         1987       6 013       1 806       7.0       5.0       35.0         1988       2 063       4 964       7.0       5.0       35.0         1990       1 886       2 332       7.0       5.0       35.0         1991       1 362       2 125       7.0       5.0       35.0         1992       2 490       2 671       7.0       5.0       35.0         1994       2 810       2 290       7.0       5.0       35.0	
1982       1 680       2 520       3.5       1.5       35.6         1983       1 800       2 700       3.5       1.5       35.6         1984       2 960       4 440       3.5       1.5       35.6         1985       1 100       3 330       3.5       1.5       35.6         1986       3 400       3 400       7.0       5.0       35.6         1987       6013       1 806       7.0       5.0       35.6         1988       2 063       4 964       7.0       5.0       35.6         1990       1 886       2 332       7.0       5.0       35.6         1991       1 862       2 125       7.0       5.0       35.6         1992       2 490       2 671       7.0       5.0       35.6         1993       3 581       1 254       7.0       5.0       35.6         1994       2 810       2 290       7.0       5.0       30.6	
1983     1 800     2 700     3.5     1.5     35.6       1984     2 960     4 440     3.5     1.5     35.6       1985     1 100     3 330     3.5     1.5     35.6       1986     3 400     7.0     5.0     35.6       1987     6 013     1 806     7.0     5.0     35.6       1988     2 063     4 964     7.0     5.0     35.6       1990     1 124     2 282     7.0     5.0     35.6       1991     1 362     2 332     7.0     5.0     35.6       1992     2 490     2 671     7.0     5.0     35.6       1993     3 581     1 254     7.0     5.0     35.6       1994     2 810     2 290     7.0     5.0     30.6	
1984     2 960     4 440     3.5     1.5     35.6       1985     1 100     3 330     3.5     1.5     35.6       1986     3 400     3 400     7.0     5.0     35.6       1987     6 013     1 806     7.0     5.0     35.6       1988     2 063     4 964     7.0     5.0     35.6       1989     1 124     2 282     7.0     5.0     35.6       1991     1 362     2 322     7.0     5.0     35.6       1992     2 490     2 671     7.0     5.0     35.6       1993     3 581     1 254     7.0     5.0     35.6       1994     2 810     2 290     7.0     5.0     30.6	
1985     1 100     3 330     3.5     1.5     35.0       1986     3 400     3 400     7.0     5.0     35.0       1987     6 013     1 806     7.0     5.0     35.0       1988     2 063     4 964     7.0     5.0     35.0       1990     1 886     2 332     7.0     5.0     35.0       1991     1 362     2 125     7.0     5.0     35.0       1992     2 490     2 671     7.0     5.0     35.0       1993     3 581     1 254     7.0     5.0     35.0       1994     2 810     2 290     7.0     5.0     30.0	
1986     3 400     3 400     7.0     5.0     35.0       1987     6 013     1 806     7.0     5.0     35.0       1988     2 063     4 964     7.0     5.0     35.0       1989     1 124     2 282     7.0     5.0     35.0       1990     1 886     2 332     7.0     5.0     35.0       1991     1 362     2 125     7.0     5.0     35.0       1992     2 490     2 671     7.0     5.0     35.0       1993     3 581     1 254     7.0     5.0     35.0       1994     2 810     2 290     7.0     5.0     30.0	
1987     6 013     1 806     7.0     5.0     35.0       1988     2 063     4 964     7.0     5.0     35.0       1989     1 124     2 282     7.0     5.0     35.0       1990     1 886     2 332     7.0     5.0     35.0       1991     1 362     2 125     7.0     5.0     35.0       1992     2 490     2 671     7.0     5.0     35.0       1993     3 581     1 254     7.0     5.0     35.0       1994     2 810     2 290     7.0     5.0     30.0	
1988     2 063     4 964     7.0     5.0     35.0       1989     1 124     2 282     7.0     5.0     35.0       1990     1 886     2 332     7.0     5.0     35.0       1991     1 362     2 125     7.0     5.0     35.0       1992     2 490     2 671     7.0     5.0     35.0       1993     3 581     1 254     7.0     5.0     35.0       1994     2 810     2 290     7.0     5.0     30.0	
1989     1 124     2 282     7.0     5.0     35.0       1990     1 886     2 332     7.0     5.0     35.0       1991     1 362     2 125     7.0     5.0     35.0       1992     2 490     2 671     7.0     5.0     35.0       1993     3 581     1 254     7.0     5.0     35.0       1994     2 810     2 290     7.0     5.0     30.0	
1990     1 886     2 332     7.0     50     35.0       1991     1 362     2 125     7.0     5.0     35.0       1992     2 490     2 671     7.0     5.0     35.0       1993     3 581     1 254     7.0     5.0     35.0       1994     2 810     2 290     7.0     5.0     30.0	
1991     1 362     2 125     7.0     5.0     35.0       1992     2 490     2 671     7.0     5.0     35.0       1993     3 581     1 254     7.0     5.0     35.0       1994     2 810     2 290     7.0     5.0     30.0	
1992     2 490     2 671     7.0     5.0     35.0       1993     3 581     1 254     7.0     5.0     35.0       1994     2 810     2 290     7.0     5.0     30.0	
1993 3 581 1 254 7.0 5.0 35.0 1994 2 810 2 290 7.0 5.0 30.0	
1994 2810 2290 7.0 5.0 30.0	
1995 1 669 1 095 12.5 7.5 30.0	10.0
1996 2 063 1 943 12.5 7.5 30.0	
1997 1060 1001 12.5 7.5 30.0	
1998 2065 846 12.5 7.5 30.0	
1999 690 1 831 12.5 7.5 30.0	
2000 1792 1277 12.5 7.5 30.0	10.0
2001 1544 1489 12.5 7.5 30.0	10.0
2002 2423 1065 30.0 10.0 22.5 7.5 12.5 7.5 30.0	
2003 1598 1540 30.0 10.0 22.5 7.5 12.5 7.5 30.0	
2004 1 927 2 880 30.0 10.0 22,5 7.5 12.5 7.5 30.0	10.0
2005 1 256 1 771 30.0 10.0 22.5 7.5 12.5 7.5 30.0	
2006 1763 1785 30.0 10.0 22.5 7.5 12.5 7.5 30.0	
2007 1378 1685 30.0 10.0 22.5 7.5 12.5 7.5 30.0	
2008 1365 1865 30.0 10.0 22.5 7.5 12.5 7.5 30.0	
2009 389 863 30.0 10.0 22.5 7.5 12.5 7.5 30.0	
2010 1313 711 30.0 10.0 22.5 7.5 12.5 7.5 30.0	
2011 899 1998 300 10.0 22.5 7.5 12.5 7.5 30.0	
2012 974 1585 30.0 10.0 22.5 7.5 12.5 7.5 30.0	
2013 1371 1632 30.0 10.0 22.5 7.5 12.5 7.5 30.0	10.0
2014 1217 2027 30.0 10.0 22.5 7.5 12.5 7.5 30.0	10.0
2015 1 1 24 2 286 30.0 10.0 22.5 7.5 12.5 7.5 30.0	10.0
2016 1 017 972 30.0 10.0 22.5 7.5 12.5 7.5 30.0	
2017 1524 986 30.0 10.0 22.5 7.5 12.5 7.5 30.0	10.0

#### Iceland (South and West)

Annual input data for NEAC PFA run-reconstruction & NCL models for ICELAND (SOUTH-WEST). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

			=		=					
Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW		rate (%) - MSW	
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	30 618	16 749	2	1	2	1	50	10	70	10
1972	24 832	25 733	2	1	2	1	50	10	70	10
1973	26 624	23 183	2	1	2	1	50	10	70	10
1974 1975	18 975 29 428	20 017 21 266	2 2	1	2 2	1 1	50 50	10 10	70 70	10 10
1975	23 233	18 379	2	1	2	1		10	70	
1976	23 233	17 919	2	1	2	1	50 50	10	70	10 10
1977	31 199	23 182	2	1	2	1	50	10	70	10
1978	28 790	14 840	2	1	2	1	50	10	70	10
1980	13 073	20 855	2	1	2	1	50	10	70	10
1980	16 890	13 919	2	1	2	1	50	10	70	10
1982	17 331	9826	2	1	2	1	50	10	70	10
1983	21 923	16 423	2	1	2	1	50	10	70	10
1984	13 476	13 923	2	1	2	1	50	10	70	10
1985	21 822	10 097	2	1	2	1	50	10	70	10
1986	35 891	8 423	2	1	2	1	50	10	70	10
1987	22 302	7 480	2	1	2	1	50	10	70	10
1988	40 028	8 523	2	1	2	1	50	10	70	10
1989	22 377	7 607	2	1	2	1	50	10	70	10
1990	20 584	7 548	2	1	2	1	50	10	70	10
1991	22 711	7 5 1 9	2	1	2	1	50	10	70	10
1992	26 006	8 479	2	1	2	1	50	10	70	10
1993	25 479	4 155	2	1	2	31	50	10	70	10
1994	20 985	6 736	2	1	2	1	50	10	70	10
1995	25 371	6777	4	1	4	1	50	10	70	10
1996	21 913	4 364	4	1	4	1	50	10	70	10
1997	16 007	4910	4	1	4	1	50	10	70	10
1998	21 900	3 037	4	1	4	1	50	10	70	10
1999	17 448	5 757	4	1	4	1	49	10	68	10
2000	15 502	1 519	4	1	4	1	49	10	66	10
2001	13 586	2 707	4	1	4	1	48	10	67	10
2002	16 952	2 845	4	1	4	1	48	10	65	10
2003	20 271	4 751	4	1	4	1	48	10	68	10
2004	20 319	3 784	4	1	4	1	48	10	67	10
2005	29 969	3 241	4	. 1	4	1	48	10	65	10
2006	21 153	2 689	4	1	4	1	48	10	65	10
2007	23 728	1 679	4	1	4	1	47	9	66	10
2008	28 774	1 659	4		4	1	47	10	57	10
2009	33 190	2 838	4	1	4	1	48	10	63	10
2010	33 318	6 0 6 1	4	1	4	1	47	10	65	10
2011	23 436	2 934	4	1	4	1	47	10	62	10
2012	13 312	1 429	4	1	4	1	47	10	53	10
2013	39 637	4 105	4	1	4	1	47	10	55	10
2014	9 551	2 281	4	1	4	1	46	10	50	10
2015	26 082	2 197	4,	1	4	1	45	10	53	10
2016	15 291	2 784	4	1	4	1	45	10	47	10
2017	16 018	4 892	4	1	4	1	46	10	61	10

### Iceland (North and East)

Annual input data for NEAC PFA run-reconstruction & NCL models for ICELAND (NORTH-EAST). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW		rate (%) - MSW	
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	4 610	6 625	2	1	2	1	50	10	70	10
1972	4 223	10 337	2	1	2	1	50	10	70	10
1973	5 060	9 672	2	1	2	1	50	10	70	10
1974 1975	5 047	9 176	2 2	1	2 2	1 1	50 50	10 10	70 70	10 10
1975	6 152 6 184	10 136 8 350	2	1	2	1		10	70	
1976	8 597	11 631	2	1	2	1	50 50	10	70	10 10
1977	8 739	14 998	2	1	2	1	50	10	70	10
1978	8 363	9 897	2	1	2	1	50	10	70	10
1980	1 268	13 784	2	1	2	1	50	10	70	10
1980	6 528	4 827	2	1	2	1	50	10	70	10
1982	3 007	5 539	2	1	2	1	50	10	70	10
1983	4 437	4 224	2	1	2	1	50	10	70	10
1984	1611	5 447	2	1	2	1	50	10	70	10
1985	11 116	3 511	2	1	2	1	50	10	70	10
1986	13 827	9 569	2	1	2	1	50	10	70	10
1987	8 145	9 908	2	1	2	1	50	10	70	10
1988	11 775	6 381	2	1	2	1	50	10	70	10
1989	6 342	5 414	2	1	2	1	50	10	70	10
1990	4 752	5 709	2	1	2	1	50	10	70	10
1991	6 900	3 965	2	1	2	1	50	10	70	10
1992	12 996	5 903	2	1	2	1	50	10	70	10
1993	10 689	6 672	2	1	2	1	50	10	70	10
1994	3 414	5 656	2	1	2	- 1	50	10	70	10
1995	8 776	3 5 1 1	4	1	4	1	50	10	70	10
1996	4 681	4 605	4	1	4	1	50	10	70	10
1997	6 406	2 594	4	1	4	1	50	10	70	10
1998	10 905	3 780	4	1	4	1	50	10	70	10
1999	5 326	4 030	4	1	4	1	48	10	65	10
2000	5 595	2 324	4	1	4	1	48	10	64	10
2001	4 976	2 587	4	1	4	1	47	10	62	10
2002	8 437	2 366	4	1	4	1	46	10	60	10
2003	4 478	2 194	4	1	4	1	46	10	53	10
2004	11 823	2 239	4	1	4	1	45	10	55	10
2005	10 297	2 726	4	. 1	4	1	44	10	54	10
2006	11 082	2 179	4	1	4	1	45	10	45	10
2007	8 046	1 672	4	1	4	1	44	10	36	10
2008	7 021	2 693	4		4	1	42	10	45	10
2009	10 779	1 735	4	1	4	1	40	10	36	10
2010	8 621	2 602	4	1	4	1	40	10	38	10
2011	6 759	2 596	4	1	4	1	38	10	34	10
2012	3 699	1 419	4	1	4	1	40	10	33	10
2013	8 375	1 528	4	1	4	1	38	10	31	10
2014	3 953	1 778	4	1	4	1	38	10	30	10
2015	10 209	1 803	4,	1	4	1	35	10	32	10
2016	4 237	2 298	4	1	4	1	34	29	34	10
2017	4 208	1 249	4	1	4	1	32	21	32	10

Ireland

Annual input data for NEAC PFA run-reconstruction & NCL models for IRELAND. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared				Estimated %			Uncertainty in				Declared net	Catch and		1SW salmon MSW salmon		MSW salmon
	catch 1SW	catch MSW		% unreported				exploitation		exploitation	catch 1SW				in Small rivers in Small rivers	in closed	in closed
	salmon	salmon	salmon	salmon	catch of MSW salmon	salmon	rate (%) - 15W salmon		rate (%) - MSW salmon	rate (%) - MSW salmon	salmon	salmon	salmon	salmon		rivers	rivers
1971	409 965	46 594	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5							
1971	437 089	49 863	37.5 37.5		37.5 37.5	7.5		12.5	47.5	12.5							
1973	476 131	54 008	37.5	7.5	37.5	7.5		12.5	47.5	12.5							
1974	542 124	60 976	37.5		37.5	7.5		12.5	47.5	12.5							
1975	598 524	68 260	37.5	7.5	37.5	7.5		12.5	47.5	12.5							
1976	407 018	47 358	37.5	7.5	37.5	7.5		12.5	47.5	12.5							
1977	351 745	41 256	37.5		37.5	7.5		12.5	47.5	12.5							
1978	307 569	35 708	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5							
1979	282 700	32 144	37.5	7.5	37.5	7.5		12.5	47.5	12.5							
1980	215 116	35 447	37.5	7.5	37.5	7.5	62.5	12.5	47.5	12.5				•			
1981	137 366	26 101	37.5		37.5	7.5		11.4	47.5	12.5		<b>- X</b>					
1982	269 847	11 754	37.5	7.5	37.5	7.5	71.9	10.8	36.7	8.3		_					
1983	437 751	26 479	37.5	7.5	37.5	7.5	66.1	9.9	40.1	7.5							
1984	224 872	20 685	37.5	7.5	37.5	7.5	64.6	9.7	43.5	6.5							
1985	430 315	18 830	37.5	7.5	37.5	7.5	74.6	11.2	36.1	3.4							
1986	443 701	27 111	37.5	7.5	37.5	7.5	68.7	10.3	46.0	9.0							
1987	324 709	26 301	30.0	10.0	30.0	10.0	69.8	10.5	32.2	4.7							
1988	391 475	22 067	30.0	10.0	30.0	10.0	62.0	9.3	37.4	5.6							
1989	297 797	25 447	30.0	10.0	30.0	10.0	65.7	9.9	47.2	8.8							
1990	172 098	15 549	30.0	10.0	30.0	10.0	60.7	9.1	59.9	6.1							
1991	120 408	10 334	30.0	10.0	30.0	10.0	59.5	8.9	26.5	3.5							
1992	182 255	15 456	30.0	10.0	30.0	10.0	62.1	9.3	51.5	3.8							
1993	150 274	13 156	25.0	10.0	25.0	10.0	58.6	8.8	42.0	18.0							
1994	234 126	20 506	25.0	10.0	25.0	10.0	71.4	10.7	40.5	2.5							
1995	232 480	20 454	25.0	10.0	25.0	10.0	63.5	9.5	41.8	1.2							
1996	203 920	18 021	25.0	10.0	25.0	10.0	59.9	9.0	55.1	3.2							
1997	170 774	14 724	25.0	10.0	15.0	5.0	50.1	7.5	30.8	12.2							
1998	191 868	17 269	25.0	10.0	15.0	5.0		8.1	61.9	1.4							
1999	158 818	14 801	25.0	10.0	15.0	5.0	47.8	7.2	34.1	18.1							
2000	199 827	16 848	25.0	10.0	15.0	5.0	43.2	6.5	31.0	4.5							
2001	218 715	18 436	7.5		7.5	2.5	48.0	7.2	35.0	8.0							
2002	198 719	16 702	7.5	2.5	7.5	2.5	49.9	7.5	27.5	7.5							
2003	161 270	13 745	7.5		7.5	2.5	41.3 49.5	6.2	21.5	5.5							
2004	142 251	12 299	7.5	2.5 2.5	7.5	2.5			35.0	8.0							
2005 2006	127 371 101 938	10 716 9 740	7.5 7.5		7.5 7.5	2.5 2.5		6.5 6.5	23.5 29.5	3.5 13.5							
											0.177	ccc	12 127	000	0 0	24.422	150
2007 2008	17 863 31 843	2 867 3 935	7.5 7.5		7.5 7.5	2.5 2.5	15.5 15.5	8.4 8.4	23.9 23.9	9.1 9.1	8 177 8 233		12 137 12 071	988 1 492	0 0		158 213
2008	24 268	4 675	7.5 7.5		7.5	2.5		8.4	23.9	9.1	6 248		9812	1 610	0 0		1873
2010	24 268 32 981	4 497	7.5	2.5	7.5	2.5		8.4	23.9	9.1	13 093		13 325	1 817	0 0		616
2010	28 105	4 497	7.5 7.5		7.5	2.5	15.5	8.4	23.9	9.1	11 071		11 031	1 657	0 0		765
2011	28 105	4 889	7.5	2.5	7.5	2.5		8.4	23.9	9.1	9 542		10 429		0 0		1 213
2012	24 029	4 831	7.5		7.5	2.5	15.5	8.4	23.9	9.1	13 378		8 821		0 0		1 250
2013	13 787	4 063	7.5	2.5	7.5	2.5		8.4	23.9	9.1	9 173		5 107	1 430	0 0		1 23
2014	20 835	4 272	7.5		7.5	2.5		8.4	23.9	9.1	7 396		7 810		0 0		1 147
2015	21 084	3 783	7.5		7.5	2.5	15.5	8.4	23.9	9.1	5 755		8 890		0 0	23 952	1 657
2017	26 023	5 267	7.5		7.5	2.5		8.4	23.9	9.1	5 892		9 415		0 0		2 511

Norway (Southeast)

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (SOUTH-EAST). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared catch 1SW salmon	Declared catch MSW salmon	Estimated % unreported catch of 1SW salmon	Uncertainty in % unreported catch of 1SW salmon	Estimated % unreported catch of MSW salmon	Uncertainty in % unreported catch of MSW salmon	Estimated exploitation rate (%) - 1SW salmon	Uncertainty in exploitation rate (%) - 1SW salmon	Estimated exploitation rate (%) - MSW salmon	Uncertainty in exploitation rate (%) - MSW salmon
1971										
1972										
1973										
1974										
1975										
1976										
1977										
1978										
1979										
1980										
1981										
1982										
1983	9 039	9 004	50	10	50	10	70	10	65	10
1984	11 402	11 527	50	10	50	10	70	10	65	10
1985	18 699	11 883	50	10	50	10	70	10	65	10
1986	23 089	12 077	50	10	50	10	70	10	65	10
1987	19 601	14 179	50	10	50	10	70	10	65	10
1988	17 520	9 443	50	10	50	10	70	10	65	10
1989	23 965	12 254	50	10	50	10	65	10	60	10
1990	25 792	11 502	50	10	50	10	65	10	60	10
1991	21 064	10 753	50	10	50	10	65	10	60	10
1992	26 044	15 332	50	10	50	10	65	10	60	10
1993	23 070	12 596	40	10	40	10	65	10	60	10
1994	23 987	9 988	40	10	40	10	65	10	60	10
1995	21 847	11 630	40	10	40	10	65	10	60	10
1996	20 738	13 538	40	10	40	10	65	10	60	10
1997	21 121	7 756	35	10	35	10	60	10	60	10
1998	32 586	10 396	35	10	35	10	60	10	60	10
1999	23 904	6 664	35	10	35	10	60	10	60	10
2000	43 151	14 261	35	10	35	10	60	10	60	10
2001	47 339	19 210	35	10	35	10	60	10	60	10
2002	33 087	14 400	35	10	35	10	60	10	60	10
2003	33 371	20 648	30	10	30	10	60	10	60	10
2004	28 506	15 948	30	10	30	10	60	10	60	10
2005	40 628	14 628	30	10	30	10	60	10	60	10
2006	30 979	21 192	30	10	30	10	60	10	60	10
2007	15 735	18 130	30	10	30	10	60	10	60	10
2008	15 696	16 678	30	10	30	10	55	10	50	10
2009	15 584	11 995	30	10	30	10	55	10	50	10
2010	22 139	12 175	30	10	30	10	50	10	40	10
2011	15 773	28 589	30	10	30	10	50	10	40	10
2012	18 582	23 389	30	10	30	10	50	10	40	10
2013	16 702	13 564	30	10	30	10	50	10	40	10
2014	15 389	13 699	30	10	30	10	40	10	35	10
2015	17 188	17 079	30	10	30	10	50	10	40	10
2016	14 670	18 579	30	10	30	10	50	10	40	10
2017	16 921	20 829	30	10	30	10	50	10	40	10

### Norway (Southwest)

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (SOUTH-WEST) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971										
1972										
1973										
1974										
1975										
1976										
1977										
1978										
1979										
1980 1981										
1982	24.045	20.001		10		10	00	10	00	10
1983	31 845	28 601	50	10	50	10	80	10	80	10
1984 1985	23 428 29 857	27 641	50 50	10	50	10 10	80 80	10 10	80 80	10
		25 515		10	50					10
1986	29 894	30 769	50	10	50	10	80	10	80	10
1987 1988	30 005	26 623	50	10	50	10 10	80	10	80	10
	36 976	28 255	50	10	50		80	10	80	10
1989	19 183	13 041	50	10	50	10	70	10	65	10
1990	18 490	14 423	50	10	50	10	70	10	65	10
1991	9 759	8 323	50	10	50	10	70 70	10 10	65	10
1992	6 448	8 832	50	10	50	10			65	10
1993	11 433	10 239	40	10	40	10	70	10	65	10
1994	18 597	10 961	40	10	40	10 10	70	10	65	10
1995	10 863	13 122	40	10	40		70	10	65	10
1996	7 048	12 546	40	10	40	10	70	10	65	10
1997	10 279	7 194	35	10	35		60	10	60	10
1998	5 726 7 357	6 583	35	10	35	10	60	10	60	10
1999		3 219	35	10	35	10	60	10	60	10
2000	11 538	7 961	35	10	35 35	10 10	60	10	60	10
2001	12 109	10 716	35	10			60	10	60	10
2002 2003	6 000 8 269	7 145 7 602	35 30	10 10	35 30	10	60 60	10	60 60	10 10
2003							60	10	60	
	7 180	6 420	30	10	30	10		10		10
2005	10 370	7 334	30	10	30	10	60	10	60	10
2006	5 173	9 381	30	10	30 30 30 30	10	60	10	60	10
2007 2008	2 630	6 011 4 807	30 30	10	30	10 10	60 55	10 10	60 50	10 10
	3 143			10						
2009 2010	3 069 3 450	3 79 <u>2</u> 2 447	30 30	10 10	30 30	10 10	55 50	10 10	50 35	10 10
2010	2 888	4 409	30	10	30	10	50 45	10	35	10
2011	4 171		30	10	30	10	45 45	10	30	10
2012	4 1/1 3 111	5 733 3 581	30	10	30	10	45 45	10	30	
2013	3 029	2 717	30	10	30	10	45	10	25	10 10
2014	4 721	3.953	30	10	30	10	40 45	10	30	10
				•					30	
2016 2017	3 262 2 009	5 671 4 547	30 30	10 10	30 30	10 10	45 45	10 10	30	10 10
2017	2 009	434/	30	10	30	10	45	10	30	10

Mid-Norway

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (MID AREA) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared catch 1SW salmon	Declared catch MSW salmon	Estimated % unreported catch of 1SW salmon	Uncertainty in % unreported catch of 1SW salmon	Estimated % unreported catch of MSW salmon	Uncertainty in % unreported catch of MSW salmon	Estimated exploitation rate (%) - 1SW salmon	Uncertainty in exploitation rate (%) - 1SW salmon	Estimated exploitation rate (%) - MSW salmon	Uncertainty in exploitation rate (%) - MSW salmon
1971										
1972										
1973										
1974										
1975										
1976										
1977										
1978										
1979										
1980										
1981										
1982	424 224	74.540		40		40	75	40	75	10
1983 1984	121 221 94 373	74 648 67 639	50 50	10 10	50 50	10 10	75 75	10 10	75 75	10 10
1984	114 613	56 641	50	10	50	10	75 75	10	75 75	10
1986	106 921	77 225	50	10	50	10	75	10	75 75	10
1987	83 669	62 216	50	10	50	10	75	10	75 75	10
1988	80 111	45 609	50	10	50	10	75	10	75	10
1989	94 897	30 862	50	10	50	10	65	10	65	10
1990	78 888	40 174	50	10	50	10	65	10	65	10
1991	67 370	30 087	50	10	50	10	65	10	65	10
1992	51 463	33 092	50	10	50	10	65	10	65	10
1993	58 326	28 184	40	10	40	10	65	10	65	10
1994	113 427	33 520	40	10	40	10	65	10	65	10
1995	57 813	42 696	40	10	40	10	65	10	65	10
1996	28 925	31 613	40	10	40	10	65	10	65	10
1997	43 127	20 565	35	10	35	10	60	10	60	10
1998	63 497	26 817	35	10	35	10	60	10	60	10
1999	60 689	28 792	35	10	35	10	60	10	60	10
2000	109 278	42 452	35	10	35	10	60	10	60	10
2001	88 096	52 031	35	10	35	10	60	10	60	10
2002	42 669	52 774	35	10	35	10	60	10	60	10
2003	91 118	46 963	30	10	30	10	60	10	60	10
2004	38 286	49 760	30	10	30	10	60	10	60	10
2005	63 749	37 941	30	10	30	10	60	10	60	10
2006	46 495	47 691	30	10	30	10	60	10	60	10
2007	26 608	33 106	30	10	30	10	60	10	60	10
2008	31 936	34 869	30	10	30	10	55	10	45	10
2009	26 267	30 715	30	10	30	10	55	10	45	10
2010	37 557	30 524	30	10	30	10	50	10	45	10
2011	20 932	37 272	30	10	30	10	50	10	45	10
2012	22 368	28 265	30	10	30	10	50	10	45	10
2013	25 121	17 727	30	10	30	10	45	10	40	10
2014	25 349	14 199	30	10	30	10	35	10	32	10
2015	30 932	30 457	30	10	30	10	45	10	40	10
2016	20 498	26 325	30	10	30	10	45	10	40	10
2017	32 496	28 555	30	10	30	10	45	10	40	10

**Norway North** 

Annual input data for NEAC PFA run-reconstruction & NCL models for NORWAY (NORTH) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

Year	Declared catch 1SW salmon	Declared catch MSW salmon	Estimated % unreported catch of 1SW	Uncertainty in % unreported catch of 1SW	Estimated % unreported catch of MSW	Uncertainty in % unreported catch of MSW	Estimated exploitation rate (%) - 1SW		Estimated exploitation rate (%) - MSW	
1071			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971 1972										
1973										
1974										
1975										
1976										
1977										
1978										
1979										
1980										
1981										
1982										
1983	104 040	49 413	50	10	50	10	80	10	80	10
1984	150 372	58 858	50	10	50	10	80	10	80	10
1985	118 841	58 956	50	10	50	10	80	10	80	10
1986	84 150	63 418	50	10	50	10	80	10	80	10
1987	72 370	34 232	50	10	50	10	80	10	80	10
1988	53 880	32 140	50	10	50	10	80	10	80	10
1989	42 010	13 934	50	10	50	10	70	10	70	10
1990	38 216	17 321	50	10	50	10	70	10	70	10
1991	42 888	21 789	50	10	50	10	70	10	70	10
1992	34 593	19 265	50	10	50	10	70	10	70	10
1993	51 440	39 014	40	10	40	10	70	10	70	10
1994	37 489	33 411	40	10	40	10	70	10	70	10
1995	36 283	26 037	40	10	40	10	70	10	70	10
1996	40 792	36 636	40	10	40	10	70	10	70	10
1997	39 930	30 115	35	10	35	10	70	10	70	10
1998	46 645	34 806	35	10	35	10	70	10	70	10
1999	46 394	46 744	35	10	35	10	70	10	70	10
2000	61 854	51 569	35	10	35	10	70	10	70	10
2001	46 331	54 023	35	10	35	10	70	10	70	10
2002	38 101	43 100	35	10	35	10	70	10	70	10
2003	44 947	35 972	30	10	30	10	70	10	70	10
2004	34 640	28 077	30	10	30	10	70	10	70	10
2005	45 530	33 334	30	10	30	10	70	10	70	10
2006	48 688	39 508	30	10	30	10	70	10	70	10
2007	28 748	44 550	30	10	30	10	70	10	70	10
2008	34 338	40 553	30	10	30	10	65	10	65	10
2009	22 511	28 241	30	10	30	10	65	10	65	10
2010	29 836	28 611	30	10	30	10	65	10	55	10
2011	26 813	27 233	30	10	30	10	65	10	55	10
2012	28 289	28 000	30	10	30	10	65	10	55	10
2013	20 021	24 689	30	10	30	10	65	10	55	10
2014	35 171	23.816	30	10	30	10	65	10	55	10
2015	25 426	23 890	30	10	30	10	65	10	55	10
2016	23 589	33 607 31 040	30 30	10	30 30	10	65	10	55 55	10
2017	29 868	31 040	30	10	30	10	65	10	55	10

### Russia (Archangelsk and Karelia)

Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (ARCHANGLE/KORELIA). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
real	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW		
	Saillioli	Samion	salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	134	16 592	10	5	10	5	60	20	60	20
1972	116	14 434	10	5	10	5	60	20	60	20
1973	169	20 924	10	5	10	5	60	20	60	20
1974	170	21 137	10	5	10	5	60	20	60	20
1975	140	17 398	10	5	10	5	60	20	60	20
1976	111	13 781	10	5	10	5	60	20	60	20
1977	78	9 722	10	5	10	5	60	20	60	20
1978	82	10 134	10	5	10	5	60	20	60	20
1979	112	13 903	10	5	10	5	60	20	60	20
1980	156	19 397	10	5	10	5	60	20	60	20
1981	68	8 394	10	5	10	5	60	20	60	20
1982	71	8 797	10	5	10	5	60	20	60	20
1983	48	11 938	10	5	10	5	60	20	60	20
1984	21	10 680	10	5	10	5	60	20	60	20
1985	454	11 183	10	5	10	5	60	20	60	20
1986	12	12 291	10	5	10	5	60	20	60	20
1987	647	8 734	10	5	10	5	60	20	60	20
1988	224	9 978	10	5	10	5	60	20	60	20
1989	989	10 245	10	5	10	5	60	20	60	20
1990	1 418	8 429	15	5	15	5	60	20 20	60	20
1991	421	8 725	20	5	20	5	60		60	20
1992	1 031	3 949	25	5	25	5	60	20	60	20
1993	196	4 251	30	5	30	5	60	20	60	20
1994	334	5 631	35	5	35	5	60	20	60	20
1995	386	5 214	45	5	45	5	60	20	60	20
1996	231	3 753	55	5	55	5	60	20	60	20
1997	721	3 351	55	5	55	5	60	20	60	20
1998	585	4 208	55	5	55	5	60	20	60	20
1999	299	3 101	55	5	55	5	60	20	60	20
2000	514	3 382	55	5	55	5	60	20	60	20
2001	363	2 348	55	5	55	5	60	20	60	20
2002	1 676	2 439	55	5	55	5	60	20	60	20
2003	893	2 041	55	5	55	5	60	20	60	20
2004	990	3 761	55	5	55	5	60	20	60	20
2005	1 349	4915	55	5	55	5	60	20	60	20
2006	2 183	2 841	55	5	55	5	60	20	60	20
2007	1 618	2 621	55	5	55	5	60	20	60	20
2008	332	2 496	55		55	5	60	20	60	20
2009	252	2 214	55	3	55	5	60	20	60	20
2010	397	3 823	55	5	55 55	5	60	20	60	20
2011	313	2 585	55	5		5	60	20	60	20
2012	1 332	2 446		5	55	5	60	20	60	20
2013	2 296	3 480	55	5	55	5	60	20	60	20
2014	2 084	3 463	55	5	55	5	60	20	60	20
2015 2016	2 071 3 042	3 542 2 221	55 55	5 5	55 55	5 5	60 60	20 20	60 60	20 20
			55	5	55	5	60	20	60	
2017	671	2 963	55	5	55	5	60	20	60	20

#### Russia (Kola Peninsula: Barents Sea Basin)

Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (KOLA-BARENT SEA). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Salmon   Salmon   Salmon   Catch of ISW   Salmon   Salm	Year	Declared catch 1SW	Declared catch MSW	Estimated % unreported	Uncertainty in % unreported	Estimated % unreported	Uncertainty in % unreported	Estimated exploitation	Uncertainty in exploitation	Estimated exploitation	Uncertainty in exploitation
1971   4 892   5 979   15   5   15   5   45.0   5.0   45.0											
1971		301111011	341111011								salmon
1972   7978   9750   15   5   15   5   45.0   5.0   45.0     1973   9376   11 460   15   5   15   5   40.0   5.0   40.0     1974   12 794   15 638   15   5   15   5   40.0   5.0   40.0     1975   13 872   13 872   15   5   15   5   5   40.0   5.0   40.0     1976   14 493   14 048   15   5   15   5   5   5   50.0   5.0     1977   7257   8253   15   5   15   5   5   50.0   5.0     1978   71 06   71 13   15   5   15   5   5   50.0   5.0     1979   6707   31 41   15   5   15   5   5   50.0   5.0     1980   6621   5216   15   5   15   5   40.0   5.0   40.0     1981   4 547   5 973   15   5   15   5   5   40.0   5.0   40.0     1982   5 159   4 738   15   5   15   5   35.0   5.0     1984   9 453   12 601   15   5   15   5   35.0   5.0     1985   6 747   7 877   15   5   15   5   35.0   5.0     1986   10 147   5 352   15   5   15   5   35.0   5.0     1987   8 560   5 149   15   5   15   5   5   40.0   5.0     1988   6 644   3 655   15   5   15   5   5   40.0   5.0     1989   13 424   6 787   15   5   15   5   5   40.0   5.0     1990   16 038   8 224   15   5   15   5   5   40.0   5.0     1991   4 550   7 568   15   5   15   5   30.0   5.0     1992   13 494   7 109   15   5   15   5   30.0   5.0     1993   8 642   5 690   15   5   15   5   30.0   5.0     1994   6 101   4 632   15   5   15   5   30.0   5.0     1995   6 318   3 693   50   10   50   50   50     1997   3 564   867   25   5   5   5   5   5   5     1996   6 815   1 701   20   5   5   5   5   5   5     1996   6 815   1 701   20   5   5   5   5   5     1997   3 564   867   25   5   5   5   5   5   5     1998   1 854   200   323   50   50   10   50   10   50   50     1999   1 4456   2478   40   5   5   5   5   5   5     1990   1 638   324   40   5   5   5   5   5   5     1990   1 638   8 244   60   10   60   10   10   50   50     1990   1 638   3 693   15   5   5   5   5   5   5     1990   1 638   1 701   20   5   5   5   5   5     1990   1 638   1 701   20   5   5   5   5   5     1990   1 638   1 701   20   5   5   5   5     1990   1 638   1 701   20   5   5	1971	4 892	5 979								5.0
1973											5.0
1974   12794   15638   15   5   15   5   40.0   5.0   40.0   1975   13872   13872   15   5   15   5   5   5.0   5.0   45.0   1976   11493   14048   15   5   5   15   5   5   5.0   5.0   5.0   5.0   1977   7   7   7   7   7   7   7   7   7											5.0
1975											5.0
1976											5.0
1978	1976	11 493	14 048	15	5	15	5	55.0	5.0		5.0
1979   6 707   3 141   15   5   15   5   40.0   5.0   40.0   1980   6621   5 216   15   5   15   5   40.0   5.0   40.0   1981   4 547   5 973   15   5   15   5   40.0   5.0   40.0   1982   5 159   4 798   15   5   15   5   35.0   5.0   35.0   1982   5 159   4 798   15   5   15   5   35.0   5.0   35.0   1983   8 504   9 943   15   5   15   5   35.0   5.0   35.0   1984   9 453   12 601   15   5   15   5   35.0   5.0   35.0   1985   6 774   7 877   15   5   15   5   35.0   5.0   35.0   1985   6 774   7 877   15   5   15   5   35.0   5.0   35.0   1987   8 560   5 149   15   5   15   5   40.0   5.0   40.0   1988   6 644   3 655   15   5   15   5   15   5   35.0   5.0   35.0   1989   13 424   6 787   15   5   15   5   15   5   40.0   5.0   40.0   1990   16038   8 234   15   5   15   5   5   40.0   5.0   40.0   1990   16038   8 234   15   5   15   5   5   30.0   5.0   30.0   1991   4550   7 568   15   5   15   5   30.0   5.0   30.0   1993   8 642   5 690   15   5   15   5   15   5   30.0   5.0   30.0   1993   8 642   5 690   15   5   15   5   15   5   30.0   5.0   30.0   1994   6 101   4 632   15   5   15   5   30.0   5.0   30.0   1995   6 318   3 693   15   5   15   5   5   30.0   5.0   30.0   1996   6 815   1 701   20   5   20   5   25.0   5.0   25.0   1997   3 564   867   25   5   5   5   5   5   5   5   5	1977	7 257	8 253	15	5	15	5	50.0	5.0	50.0	5.0
1980   6 621   5 216   15   5   15   5   40.0   5.0   40.0     1981   4 547   5 973   15   5   15   5   40.0   5.0   40.0     1982   5 159   4 798   15   5   15   5   35.0   5.0   35.0     1983   8 504   9 943   15   5   15   5   35.0   5.0   35.0     1984   9 453   12 601   15   5   15   5   35.0   5.0   35.0     1985   6 774   7 877   15   5   15   5   35.0   5.0   35.0     1986   10 147   5 352   15   5   15   5   35.0   5.0   35.0     1987   8 560   5 149   15   5   15   5   40.0   5.0   40.0     1988   6 644   3 655   15   5   15   5   35.0   5.0   35.0     1989   13 424   6 787   15   5   15   5   35.0   5.0   35.0     1999   13 424   6 787   15   5   15   5   40.0   5.0   40.0     1991   4 550   7 568   15   5   15   5   30.0   5.0   30.0     1992   11 394   7 109   15   5   15   5   30.0   5.0   30.0     1993   8 642   5 690   15   5   15   5   30.0   5.0   30.0     1994   6 101   4 632   15   5   15   5   30.0   5.0   30.0     1995   6 318   3 693   15   5   15   5   5   30.0   5.0   30.0     1997   3 564   867   25   5   5   5   5   5   5   5   5     1999   1 510   424   40   5   40   5   40   5   7.5     1999   1 510   424   40   5   40   5   5   5     1000   805   323   50   50   10   50   10     2001   1095   850   406   10   60   10   10.0   5.0   20.0     2003   1 938   1 095   50   10   50   10   10.0   5.0   20.0     2004   1 095   850   50   10   60   10   10.0   5.0   20.0     2006   1 372   844   60   10   60   10   10.0   5.0   20.0     2011   3 861   1 407   60   10   60   10   20.0   5.0   25.0     2011   3 989   904   60   10   60   10   20.0   5.0   25.0     2015   727   499   60   10   60   10   60   10   20.0   5.0   25.0     2015   727   499   60   10   60   10   60   10   20.0   5.0     25.0   25.0   25.0     2015   727   499   60   10   60   10   20.0   5.0   25.0     2015   2015   727   499   60   10   60   10   20.0   5.0   25.0     2015   727   499   60   10   60   10   60   10   20.0   5.0     25.0   25.0   25.0   25.0     25.0   25.0   25.0   25.0     25.0   25.0   25.0   25	1978	7 106	7 113	15	5	15	5	55.0	5.0	55.0	5.0
1981	1979	6 707	3 141	15		15		40.0	5.0		5.0
1982   5   159   4   798   15   5   15   5   35.0   5.0   35.0   1983   8   504   9   943   15   5   15   5   35.0   5.0   35.0   1984   9   453   12   601   15   5   5   15   5   35.0   5.0   35.0   1985   6   74   78   77   15   5   15   5   35.0   5.0   35.0   1986   10   147   5   352   15   5   15   5   5   40.0   5.0   40.0   1987   8   560   5   149   15   5   15   5   5   40.0   5.0   40.0   1988   6   644   3   655   15   5   15   5   5   40.0   5.0   40.0   1989   13   424   6   787   15   5   15   5   5   40.0   5.0   40.0   1990   16   6038   8   234   15   5   15   5   5   5   40.0   5.0   40.0   1991   4   550   7   568   15   5   15   5   5   30.0   5.0   30.0   1992   13   394   7   109   15   5   5   15   5   30.0   5.0   30.0   1994   6   101   4   632   15   5   15   5   5   30.0   5.0   30.0   1994   6   101   4   632   15   5   5   15   5   30.0   5.0   30.0   1996   6   815   1   701   20   5   20   5   25.0   5.0   25.0   1998   15   6   6   6   6   6   6   6   6   6	1980	6 621	5 216	15		15		40.0			5.0
1983         8 504         9 943         15         5         15         5         35.0         5.0         35.0           1984         9 453         12 601         15         5         15         5         35.0         5.0         35.0           1986         10 147         5 352         15         5         15         5         35.0         5.0         40.0           1987         8 560         5 149         15         5         15         5         40.0         5.0         40.0           1988         6644         3 655         15         5         15         5         40.0         5.0         40.0           1989         13 424         6 787         15         5         15         5         35.0         5.0         30.0           1990         16 038         8 234         15         5         15         5         40.0         5.0         40.0           1991         4 550         7 568         15         5         15         5         30.0         5.0         30.0           1993         8 642         5690         15         5         15         5         30.0         5.0	1981	4 547	5 973	15	5	15		40.0	5.0	40.0	5.0
1984         9 453         12 601         15         5         15         5         35.0         5.0         35.0           1986         6774         7 877         15         5         15         5         35.0         5.0         35.0           1987         8 560         5 149         15         5         15         5         40.0         5.0         40.0           1988         6 644         3 655         15         5         15         5         40.0         5.0         40.0           1990         16 038         8 234         15         5         15         5         40.0         5.0         40.0           1991         4 550         7 568         15         5         15         5         40.0         5.0         40.0           1992         11 394         7 109         15         5         15         5         30.0         5.0         30.0           1993         8 642         5 690         15         5         15         5         30.0         5.0         30.0           1994         6 101         4 632         15         5         15         5         30.0         5.0	1982	5 159	4 798	15	5	15	5	35.0	5.0	35.0	5.0
1985         6 774         7 877         15         5         15         5         35.0         5.0         35.0           1986         10 147         5 352         15         5         15         5         40.0         5.0         40.0           1987         8 560         5 149         15         5         15         5         40.0         5.0         40.0           1988         6 644         3 655         15         5         15         5         40.0         5.0         40.0           1990         16 038         8 234         15         5         15         5         40.0         5.0         40.0           1991         4 550         7 568         15         5         15         5         40.0         5.0         30.0           1992         11 394         7 109         15         5         15         5         30.0         5.0         30.0           1993         8 642         5 690         15         5         15         5         30.0         5.0         30.0           1994         6 101         4 632         15         5         15         5         30.0         5.0	1983	8 504	9 943	15	5	15	5	35.0	5.0	35.0	5.0
1986         10 147         5 352         15         5         15         5         40.0         5.0         40.0           1987         8 560         5 149         15         5         15         5         40.0         5.0         40.0           1988         6 644         3 655         15         5         15         5         35.0         5.0         30.0           1989         13 424         6 787         15         5         15         5         40.0         5.0         40.0           1990         16 638         8 224         15         5         15         5         40.0         5.0         40.0           1991         4 550         7 568         15         5         15         5         30.0         5.0         30.0           1991         4 550         7 568         15         5         15         5         30.0         5.0         30.0           1991         4 550         7 568         15         5         15         5         30.0         5.0         30.0           1992         1 1394         7 109         15         5         15         5         30.0         5.0	1984	9 453	12 601	15	5	15	5	35.0	5.0	35.0	5.0
1987         8 560         5 149         15         5         15         5         40.0         5.0         40.0           1988         6 644         3 655         15         5         15         5         35.0         5.0         35.0           1989         13 424         6 787         15         5         15         5         40.0         5.0         40.0           1991         16038         8 234         15         5         15         5         40.0         5.0         40.0           1991         4 550         7 568         15         5         15         5         30.0         5.0         30.0           1992         11 394         7 109         15         5         15         5         30.0         5.0         30.0           1993         8 642         5 690         15         5         15         5         30.0         5.0         30.0           1994         6 101         4 632         15         5         15         5         30.0         5.0         30.0           1995         6 313         3 693         15         5         15         5         30.0         5.0	1985	6774	7 877	15	5	15	5	35.0	5.0	35.0	5.0
1988         6 644         3 655         15         5         15         5         35.0         35.0         35.0           1989         13 424         6 787         15         5         15         5         40.0         5.0         40.0           1990         16 038         8 234         15         5         15         5         40.0         5.0         30.0           1991         4 550         7 568         15         5         15         5         30.0         5.0         30.0           1992         11 394         7 109         15         5         15         5         30.0         5.0         30.0           1993         8 642         5 690         15         5         15         5         30.0         5.0         30.0           1994         6 101         4 632         15         5         15         5         30.0         5.0         30.0           1995         6 318         3 693         15         5         15         5         30.0         5.0         30.0           1997         3 564         867         25         5         20         5         25.0         5.0	1986	10 147	5 352	15	5	15	5	40.0	5.0	40.0	5.0
1989	1987	8 560	5 149	15	5	15	5	40.0	5.0	40.0	5.0
1990         16 038         8 234         15         5         15         5         40.0         5.0         40.0           1991         4 550         7 568         15         5         15         5         30.0         5.0         30.0           1992         11 394         7 109         15         5         15         5         30.0         5.0         30.0           1993         8 642         5 690         15         5         15         5         30.0         5.0         30.0           1994         6 101         4 632         15         5         15         5         30.0         5.0         30.0           1995         6 318         3 693         15         5         15         5         30.0         5.0         30.0           1996         6 815         1 701         20         5         20         5         25.0         5         25         5         25         5         25.0         5         25.0         5         25.0         5         25.0         5         25.0         5         25.0         5         25.0         5         15.0         15.0         15.0         15.0         1	1988	6 644	3 655	15	5	15	5	35.0	5.0	35.0	5.0
1991         4550         7568         15         5         15         5         30.0         5.0         30.0           1992         11 394         7 109         15         5         15         5         30.0         5.0         30.0           1993         8 642         5690         15         5         15         5         30.0         5.0         30.0           1994         6 101         4 632         15         5         15         5         30.0         5.0         30.0           1996         6 318         3 693         15         5         15         5         30.0         5.0         25.0           1997         3 564         867         25         5         25         5         15.0         5.0         15.0           1998         1 854         280         35         5         35         5         15.0         5.0         15.0           1999         1 510         424         40         5         40         5         7.5         2.5         7.5         12.5         12.5         7.5         12.5         7.5         12.5         7.5         12.5         7.5         12.5	1989	13 424	6 787	15	5	15		40.0		40.0	5.0
1992         11 394         7 109         15         5         15         5         30.0         5.0         30.0           1993         8 642         5 690         15         5         15         5         30.0         5.0         30.0           1994         6 101         4 632         15         5         15         5         30.0         5.0         30.0           1995         6 318         3 693         15         5         15         5         30.0         5.0         30.0           1996         6 815         1 701         20         5         20         5         25.0         5.0         25.0           1997         3 564         867         25         5         25         5         15.0         5.0         15.0           1998         1 854         280         35         5         35         5         12.5         2.5         12.5	1990	16 038	8 234	15	5	15	5	40.0		40.0	5.0
1993         8 642         5 690         15         5         15         5         30.0         5.0         30.0           1994         6 101         4 632         15         5         15         5         30.0         5.0         30.0           1995         6 318         3 693         15         5         15         5         30.0         5.0         30.0           1996         6 815         1 701         20         5         20         5         25.0         5.0         25.0           1997         3 564         867         25         5         25         5         15.0         5.0         15.0           1998         1 884         280         35         5         35         5         12.5         2.5         12.5         12.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5         7.5         2.5	1991	4 550	7 568	15	5	15		30.0	5.0	30.0	5.0
1994         6 101         4 632         15         5         15         5         30.0         5.0         30.0           1995         6 318         3 693         15         5         15         5         30.0         5.0         30.0           1996         6 815         1 701         20         5         20         5         25.0         5.0         25.0           1997         3 564         867         25         5         25         5         15.0         5.0         15.0           1998         1 854         280         35         5         35         5         12.5         2.5         12.5           1999         1 510         424         40         5         40         5         7.5         2.5         7.5           2000         805         323         50         5         60         5         3.5         1.5         3.5           2001         591         241         60         5         60         5         3.5         1.5         3.5           2002         1 436         2478         50         10         50         10         10.0         5.0         20.0	1992	11 394	7 109	15	5	15	5	30.0	5.0	30.0	5.0
1995         6 318         3 693         15         5         15         5         30.0         5.0         30.0           1996         6 815         1 701         20         5         20         5         25.0         5.0         25.0           1997         3 564         867         25         5         25         5         15.0         5.0         15.0           1998         1 854         280         35         5         35         5         12.5         2.5         12.5           1999         1 510         424         40         5         40         5         7.5         2.5         7.5           2000         805         323         50         5         50         5         6.0         2.0         6.0           2001         591         241         60         5         60         5         3.5         1.5         3.5           2002         1 436         2 478         50         10         50         10         10.0         5.0         20.0           2003         1 938         1 095         50         10         50         10         10.0         5.0         20.0	1993	8 642	5 690	15		15	5		5.0	30.0	5.0
1996         6 815         1 701         20         5         20         5         25.0         5.0         25.0           1997         3 564         867         25         5         25         5         19.0         5.0         15.0           1998         1 854         280         35         5         35         5         12.5         2.5         12.5           1999         1 510         424         40         5         40         5         7.5         2.5         7.5           2000         805         323         50         5         60         5         6.0         2.0         6.0           2001         591         241         60         5         60         5         3.5         1.5         3.5           2002         1 436         2478         50         10         50         10         10.0         5.0         20.0           2003         1 938         1 095         50         10         50         10         10.0         5.0         20.0           2004         1 095         850         50         10         50         10         10.0         5.0         20.0	1994	6 101	4 632	15		15		30.0	5.0	30.0	5.0
1997         3 564         867         25         5         25         5         15.0         5.0         15.0           1998         1 854         280         35         5         35         5         12.5         2.5         12.5           1999         1 510         424         40         5         40         5         7.5         2.5         7.5           2000         805         323         50         5         50         5         6.0         2.0         6.0           2001         591         241         60         5         60         5         3.5         1.5         3.5           2002         1 436         2 478         50         10         50         10         10.0         5.0         20.0           2003         1 938         1 095         50         10         50         10         10.0         5.0         20.0           2004         1 095         850         50         10         50         10         10.0         5.0         20.0           2005         899         426         60         10         60         10         10.0         5.0         20.0	1995	6 318	3 693					30.0	5.0	30.0	5.0
1998         1 854         280         35         5         35         5         12.5         2.5         12.5           1999         1 510         424         40         5         40         5         7.5         2.5         7.5           2000         805         323         50         5         50         5         6.0         2.0         6.0           2001         591         241         60         5         60         -5         3.5         1.5         3.5           2002         1 436         2478         50         10         50         10         10.0         5.0         20.0           2003         1 938         1 095         50         10         50         10         10.0         5.0         20.0           2004         1 095         850         50         10         50         10         10.0         5.0         20.0           2005         859         426         60         10         60         10         10.0         5.0         20.0           2007         784         707         60         10         60         10         10.0         5.0         20.0	1996	6 815	1 701	20	5	20	5	25.0	5.0	25.0	5.0
1999         1 510         424         40         5         40         5         7.5         2.5         7.5           2000         805         323         50         5         50         5         6.0         20         6.0           2001         591         241         60         5         60         5         3.5         1.5         3.5           2002         1 436         2 2478         50         10         50         10         10.0         5.0         20.0           2003         1 938         1 095         50         10         50         10         10.0         5.0         20.0           2004         1 095         850         50         10         50         10         10.0         5.0         20.0           2005         859         426         60         10         60         10         10.0         5.0         20.0           2006         1 372         844         60         10         60         10         10.0         5.0         20.0           2007         784         707         60         10         60         10         10.0         5.0         20.0											5.0
2000         805         323         50         5         50         5         6.0         2.0         6.0           2001         591         241         60         5         60         5         3.5         1.5         3.5           2002         1 436         2478         50         10         50         10         10.0         5.0         20.0           2003         1 938         1 095         50         10         50         10         10.0         5.0         20.0           2004         1 095         850         50         10         50         10         10.0         5.0         20.0           2005         859         426         60         10         60         10         10.0         5.0         20.0           2006         1 372         844         60         10         60         10         10.0         5.0         20.0           2007         784         707         60         10         60         10         10.0         5.0         20.0           2008         1 446         997         60         10         60         10         15.0         5.0         20.0 <td></td> <td>2.5</td>											2.5
2001         591         241         60         5         60         5         3.5         1.5         3.5           2002         1 436         2478         50         10         50         10         10.0         5.0         20.0           2003         1 938         1 095         50         10         50         10         10.0         5.0         20.0           2004         1 095         850         50         10         50         10         10.0         5.0         20.0           2005         859         426         60         10         60         10         10.0         5.0         20.0           2006         1 372         844         60         10         60         10         10.0         5.0         20.0           2007         784         707         60         10         60         10         10.0         5.0         20.0           2008         1 446         997         60         10         60         10         15.0         5.0         20.0           2010         3 884         1 486         60         10         60         10         15.0         5.0 <td< td=""><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td>2.5</td></td<>											2.5
2002         1 436         2 478         50         10         50         10         10.0         5.0         20.0           2003         1 938         1 095         50         10         50         10         10.0         5.0         20.0           2004         1 095         850         50         10         50         10         10.0         5.0         20.0           2005         859         426         60         10         60         10         10.0         5.0         20.0           2006         1 372         844         60         10         60         10         10.0         5.0         20.0           2007         784         707         60         10         60         10         10.0         5.0         20.0           2008         1 446         997         60         10         60         10         15.0         5.0         20.0           2009         2 882         1 080         60         10         60         10         15.0         5.0         20.0           2010         3 884         1 486         60         10         60         10         20.0         5.0							5				2.0
2003         1 938         1 095         50         10         50         10         10.0         5.0         20.0           2004         1 095         850         50         10         50         10         10.0         5.0         20.0           2005         859         426         60         10         60         10         10.0         5.0         20.0           2006         1 372         844         60         10         60         10         10.0         5.0         20.0           2007         784         707         60         10         60         10         10.0         5.0         20.0           2008         1 446         997         60         10         60         10         15.0         5.0         20.0           2009         2 882         1 080         60         10         60         10         15.0         5.0         20.0           2010         3 884         1 486         60         10         60         10         20.0         5.0         25.0           2012         2 708         1 027         60         10         60         10         20.0         5.0							5				1.5
2004         1 095         850         50         10         50         10         10.0         5.0         20.0           2005         859         426         60         10         60         10         10.0         5.0         20.0           2006         1 372         844         60         10         60         10         10.0         5.0         20.0           2007         784         707         60         10         80         10         10.0         5.0         20.0           2008         1 446         997         60         10         60         10         15.0         5.0         20.0           2009         2 882         1 080         60         10         60         10         15.0         5.0         20.0           2010         3 884         1 486         60         10         60         10         20.0         5.0         25.0           2011         3 861         1 407         80         10         60         10         20.0         5.0         25.0           2012         2 708         1 027         60         10         60         10         20.0         5.0											5.0
2005         859         426         60         10         60         10         10.0         5.0         20.0           2006         1 372         844         60         10         60         10         10.0         5.0         20.0           2007         784         707         60         10         60         10         15.0         5.0         20.0           2008         1 446         997         60         10         60         10         15.0         5.0         20.0           2009         2 882         1 080         80         10         60         10         15.0         5.0         20.0           2010         3 884         1 486         60         10         60         10         20.0         5.0         25.0           2011         3 861         1 407         60         10         60         10         20.0         5.0         25.0           2012         2 708         1 027         60         10         60         10         20.0         5.0         25.0           2013         939         904         60         10         60         10         20.0         5.0											5.0
2006         1 372         844         60         10         60         10         10.0         5.0         20.0           2007         784         707         60         10         60         10         10.0         5.0         20.0           2008         1 446         997         60         10         60         10         15.0         5.0         20.0           2009         2 882         1 080         60         10         60         10         15.0         5.0         20.0           2010         3 884         1 486         60         10         60         10         20.0         5.0         25.0           2011         3 861         1 407         60         10         60         10         20.0         5.0         25.0           2012         2 708         1 027         60         10         60         10         20.0         5.0         25.0           2013         939         904         60         10         60         10         20.0         5.0         25.0           2014         969         289         60         10         60         10         20.0         5.0											5.0
2007         784         707         60         10         60         10         10.0         5.0         20.0           2008         1 446         997         60         10         60         10         15.0         5.0         20.0           2009         2 882         1 080         60         10         60         10         15.0         5.0         20.0           2010         3 884         1 486         60         10         60         10         20.0         5.0         25.0           2011         3 861         1 407         60         10         60         10         20.0         5.0         25.0           2012         2 708         1 027         60         10         60         10         20.0         5.0         25.0           2013         939         904         60         10         60         10         20.0         5.0         25.0           2014         969         789         60         10         60         10         20.0         5.0         25.0           2015         727         494         60         10         60         10         20.0         5.0											5.0
2008         1 446         997         60         10         60         10         15.0         5.0         20.0           2009         2 882         1 080         60         10         60         10         15.0         5.0         20.0           2010         3 884         1 486         60         10         60         10         20.0         5.0         25.0           2011         3 861         1 407         60         10         60         10         20.0         5.0         25.0           2012         2 708         1 027         60         10         60         10         20.0         5.0         25.0           2013         939         904         60         10         60         10         20.0         5.0         25.0           2014         969         789         60         10         60         10         20.0         5.0         25.0           2015         727         494         60         10         60         10         20.0         5.0         25.0											5.0
2009         2 882         1 080         60         10         60         10         15.0         5.0         20.0           2010         3 884         1 486         60         10         60         10         20.0         5.0         25.0           2011         3 861         1 407         60         10         60         10         20.0         5.0         25.0           2012         2 708         1 027         60         10         60         10         20.0         5.0         25.0           2013         939         904         60         10         60         10         20.0         5.0         25.0           2014         969         788         60         10         60         10         20.0         5.0         25.0           2015         727         494         60         10         60         10         20.0         5.0         25.0											5.0
2010     3 884     1 486     60     10     60     10     20.0     5.0     25.0       2011     3 861     1 407     60     10     60     10     20.0     5.0     25.0       2012     2 708     1 027     60     10     60     10     20.0     5.0     25.0       2013     939     904     60     10     60     10     20.0     5.0     25.0       2014     969     788     60     10     60     10     20.0     5.0     25.0       2015     727     494     60     10     60     10     20.0     5.0     25.0											5.0
2011     3 861     1 407     60     10     60     10     20.0     5.0     25.0       2012     2 708     1 027     60     10     60     10     20.0     5.0     25.0       2013     939     904     60     10     60     10     20.0     5.0     25.0       2014     969     789     60     10     60     10     20.0     5.0     25.0       2015     727     494     60     10     60     10     20.0     5.0     25.0											5.0
2012     2 708     1 027     60     10     60     10     20.0     5.0     25.0       2013     939     904     60     10     60     10     20.0     5.0     25.0       2014     969     789     60     10     60     10     20.0     5.0     25.0       2015     727     494     60     10     60     10     20.0     5.0     25.0						*					5.0
2013     939     904     60     10     60     10     20.0     5.0     25.0       2014     969     789     60     10     60     10     20.0     5.0     25.0       2015     727     494     60     10     60     10     20.0     5.0     25.0				_							5.0
2014 969 <b>789</b> 60 10 60 10 20.0 5.0 25.0 2015 727 494 60 10 60 10 20.0 5.0 25.0											5.0
2015 727 494 60 10 60 10 20.0 5.0 25.0											5.0
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											5.0 5.0
2017 265 503 60 10 60 10 20.0 5.0 25.0											5.0

### Russia (Kola Peninsula: White Sea Basin)

Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (KOLA-WHITE SEA). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Declared	Catch 1SW	Catch MSW	Estimated %		Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW		following-year		unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	spawners	spawners	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW		rate (%) - MSW	
					salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	67 845	29 077	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1972	45 837	19 644	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1973	68 684	29 436	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1974	63 892	27 382	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1975	109 038	46 730	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1976	76 281	41 075	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1977	47 943	32 392	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1978	49 291	17 307	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1979	69 511	21 369	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1980	46 037	23 241	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1981	40 172	12 747	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1982	32 619	14 840	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1983	54 217	20 840	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1984	56 786	16 893	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1985	87 274	16 876	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1986	72 102	17 681	0	0	3	2	3.0	2.0	50.0	10.0	60.0	10.0
1987	79 639	12 501	0	0	3	2	3.0	2.0	50.0	10.0	50.0	10.0
1988	44 813	18 777	0	0	3	2	3.0	2.0	45.0	5.0	45.0	5.0
1989	53 293	11 448					7.5	2.5	45.0	5.0	45.0	5.0
1990	44 409	11 152	0	0	13	3	12.5	2.5	45.0	5.0	45.0	5.0
1991	31 978	6 263	0	0	18	3	17.5	2.5	35.0	5.0	35.0	5.0
1992	23 827	3 680	0	0	23	3 5	22.5	2.5	25.0	5.0	25.0	5.0
1993	20 987	5 552	0	0	25	5	25.0	5.0	25.0	5.0	25.0	5.0
1994 1995	25 178 19 381	3 680 2 847	0	0	30 35	5	30.0 35.0	5.0	25.0 25.0	5.0 5.0	15.0 15.0	5.0 5.0
			-	0		5		5.0				
1996	27 097	2 710	0		35 35		35.0	5.0 5.0		5.0	15.0	5.0
1997	27 695	2 085	0	0	35	5 5	35.0		25.0 25.0	5.0	15.0	5.0
1998	32 693	1963	0				35.0	5.0		5.0	15.0	5.0
1999 2000	22 330	2 841	0	0	35 35	5 5	35.0 35.0	5.0	25.0 25.0	5.0	15.0	5.0
	26 376	4 396	0	0		5		5.0		5.0	15.0	5.0
2001	20 483	3 959	0	0	35 35	5	35.0 35.0	5.0	15.0 15.0	5.0	15.0	5.0
2002	19 174	3 937	0	0		5		5.0		5.0	15.0	5.0
2003 2004	15 687 10 947	3 734 1 990	0	0	35 35	5	25.0 35.0	5.0 5.0	15.0	5.0	15.0	5.0
									15.0	5.0	15.0	5.0
2005	13 172 15 004	2 388	1212	878 399	35 35	5	35.0	5.0	15.0 15.0	5.0 5.0	15.0 15.0	5.0 5.0
2006 2007	7 807	2 071 1 404	3852 2264	852	35	5 5	35.0		15.0	5.0	15.0	5.0
2007	8 447	4 711	3175	832	35	5	35.0 35.0	5.0 5.0	15.0	5.0	15.0	5.0
								_				
2009 2010	5 351 6 731	3 105 4 158	5130 3684	1710 1228	35 35	5 5	35.0 35.0	5.0 5.0	15.0 15.0	5.0 5.0	15.0 15.0	5.0 5.0
2010	7 363	4 325	3082	1027	35		35.0	5.0	15.0			
2011	10 398	4 325 1 431	3082 2267	756	35	5	35.0 35.0	5.0	15.0	5.0 5.0	15.0 15.0	5.0 5.0
							35.0 35.0					
2013 2014	8 986 8 593	1 660 1 674	2203 3307	734 1102	35 35	5	35.0	5.0 5.0	15.0 15.0	5.0 5.0	15.0 15.0	5.0 5.0
2014		1 179	3307 2964	1102	± 35	5 5	35.0 35.0					5.0
2015	9 115 5 969	1 1 / 9 848	2964 1526	626	35	5	35.0 35.0	5.0 5.0	15.0 15.0	5.0 5.0	15.0 15.0	5.0
2016	1 861	848 294	1526	531	35 35	2	35.0 35.0	5.0	15.0	5.0	15.0	5.0

### Russia (Pechora River)

Annual input data for NEAC PFA run-reconstruction & NCL models for RUSSIA (PECHORA RIVER). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Declared	Return	Return		Uncertainty in		Uncertainty in %	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW		estimate MSW		% unreported	unreported	unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	rate (%) - MSW
1971	605	17 728			salmon 20	salmon 10	salmon 20	salmon 10	salmon 65	salmon 15	salmon 65	salmon 15
1971	825	24 175			20	10	20	10	65	15	65	15
1972	1 705	49 962			20	10	20	10	65	15	65	15
1974	1 320	38 680			20	10	20	10	65	15	65	15
1975	1 298	38 046			20	10	20	10	65	15	65	15
1976	991	34 394			20	10	20	10	65	15	65	15
1977	589	20 464			20	10	20	10	65	15	65	15
1978	759	26 341			20	10	20	10	65	15	65	15
1979	421	14 614			20	10	20	10	65	15	65	15
1980	1 123	39 001			20	10	20	10	65	15	65	15
1981	126	20 874			20	10	20	10	65	15	65	15
1982	54	13 546			20	10	20	10	65	15	65	15
1983	598	16 002			20	10	20	10	65	15	65	15
1984	1 833	15 967			20	10	20	10	65	15	65	15
1985	2 763	29 738			20	10	20	10	65	15	65	15
1986	66	32 734			20	10	20	10	65	15	65	15
1987	21	21 179			20	10	20	10	65	15	65	15
1988	3 184	12 816			20	10	20	10	65	15	65	15
1989			24 596	27 404	10	5	10	5	65	15	65	15
1990			50	49 950	10	5	10	5	65	15	65	15
1991			7 975	47 025	10	5	10	5	65	15	65	15
1992			550	54 450	10	5	10	5	65	15	65	15
1993			68	67 932	10	5	10	5	65	15	65	15
1994			3 900	48 100	10	5	10	5	65	15	65	15
1995			9 280	70 720	10	5	10	5	65	15	65	15
1996			8 664	48 336	10	5	10	5	65	15	65	15
1997			1 440	38 560	10	5	10	5∉	65	15	65	15
1998			780	59 220	10	5	10	5	65	15	65	15
1999			2 120	37 880	10	5	10	5	65	15	65	15
2000			84	83 916	10	5	10	5	65	15	65	15
2001			2 244	41 756	10	5	10	5	65.	15	65	15
2002			405	44 595	10	5	10	5	65	15	65	15
2003			1 650	31 350	10	5	10	5	65	15	65	15
2004			6 075	20 925	10	5	10	5	65	15	65	15
2005			2 852	28 148	10	5	10	5	65	15	65	15
2006			1 472	30 528	10	5	10	5	65	15	65	15
2007			817	42 183	10	5	10	3	65	15	65	15
2008			300	49 700	10	5	10	5	65	15	65	15
2009			1 116	47 384	10	5	10	5	65	15	65	15
2010 2011			1 096	53 704	10	5 5	10	5 5	65	15 15	65	15
			2 990	56 810	10		10	5 5	65		65	15
2012 2013			4 424 4 225	27 176	10 10	5	10 10	5	65	15 15	65	15 15
2013			4 225 2 251	30 983 31 349	10	5	10	5	65 65	15	65 65	15
						,	_	5		15		
2015 2016			4 626 4 260	34 574 31 840	10 10		10 10	5	65 65	15	65 65	15 15
2016			4 257	31 840	10		10	5	65	15	65	15

Sweden

Annual input data for NEAC PFA run-reconstruction & NCL models for SWEDEN. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

			=		=					
Year	Declared	Declared	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW		
1971	6 220	254	salmon 30.0	salmon 15.0	salmon 30.0	salmon 15.0	salmon 52.5	salmon 12.5	salmon 57.5	salmon 12.5
1971	4 943	254	30.0	15.0	30.0	15.0	52.5	12.5	57.5 57.5	12.5
1972	6 124	895	30.0	15.0	30.0	15.0	52.5	12.5	57.5 57.5	12.5
1973	8 870	563	30.0	15.0	30.0	15.0	52.5	12.5	57.5 57.5	12.5
1974	9 620	160	30.0	15.0	30.0	15.0	52.5 52.5	12.5	57.5 57.5	12.5
1976	5 420	480	30.0	15.0	30.0	15.0	52.5	12.5	57.5	12.5
1977	2 453	206	30.0	15.0	30.0	15.0	52.5	12.5	57.5	12.5
1978	2 903	254	30.0	15.0	30.0	15.0	52.5	12.5	57.5	12.5
1979	2 903	661	30.0	15.0	30.0	15.0	52.5	12.5	57.5 57.5	12.5
1979	3 842	1 283	30.0	15.0	30.0	15.0	52.5	12.5	57.5 57.5	12.5
1980	7 013	284	30.0	15.0	30.0	15.0	52.5 52.5	12.5	57.5 57.5	12.5
1982	6 177	1 381	30.0	15.0	30.0	15.0	52.5	12.5	57.5	12.5
1983	8 222	903	30.0	15.0	30.0	15.0	52.5	12.5	57.5	12.5
1984	11 584	1 266	30.0	15.0	30.0	15.0	52.5	12.5	57.5	12.5
1985	13 810	470	30.0							
1985	14 415	240		15.0	30.0	15.0	52.5	12.5	57.5	12.5 12.5
1986	11 450	1 084	30.0 30.0	15.0 15.0	30.0 30.0	15.0 15.0	52.5 52.5	12.5 12.5	57.5 57.5	12.5
1988	9 604	1 160	30.0	15.0	30.0	15.0	52.5 52.5	12.5	57.5	12.5
1989	2 803	4 044	30.0	15.0	30.0	15.0	52.5	12.5	57.5	12.5
1990	6 839	2 249	15.0	10.0	15.0	10.0	45.0	15.0	50.0	15.0
1991	8 599	3 033	15.0	10.0	15.0		45.0	15.0		15.0
1991	9 550	4 205	15.0	10.0	15.0	10.0 10.0	45.0 45.0	15.0 15.0	50.0 50.0	15.0
1993	9 468	4 762	15.0	10.0	15.0	10.0	45.0	15.0	50.0	15.0
1994	7 347	3 628	15.0	10.0	15.0	10.0	45.0	15.0	50.0	15.0
1995	8 933	1 528	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
1996	5 318	2 507	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
1997	2 415	1809	15.0	10.0	15.0	10.0	37.5 37.5	12.5	42.5	12.5
1998	1 953	1000	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
1999	3 075	712	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
2000	5 660	2 546	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
2001	3 504	3 026	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
2002	3 374	2 075	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
2003	1 833	496	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
2003	1537	1 528	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
2005	1 503	1 027	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
2006	1 676	1 069	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
2007	521	1 003	15.0	10.0	15.0	10.0	37.5	12.5	42.5	12.5
2008	615	1 112	12.5	7.5	12.5	7.5	27.5	12.5	32.5	12.5
2009	651	979	12.5	7.5	12.5	7.5	27.5	12.5	32.5	12.5
2010	1 111	1 139	12.5	7.5	12.5	7.5	27.5	12.5	32.5	12.5
2011	1 460	3 100	17.5	7.5	17.5	7.5	45.0	15.0	50.0	15.0
2012	1 336	3 130	12.5	7.5	10.0	5.0	27.5	12.5	32.5	12.5
2013	874	1 431	10.0	5.0	10.0	5.0	30.0	15.0	35.0	15.0
2013	2 347	2 797	12.5	7.5	12.5	7.5	30.0	12.5	35.0	12.5
2015	1 028	2 569	12.5	7.5	12.5	7.5	30.0	12.5	30.0	12.5
2016	554	910	10.0	7.5	10.0	7.5	25.0	10.0	25.0	10.0
2017	529	934	10.0	7.5	10.0	7.5	25.0	10.0	25.0	10.0

UK (England and Wales)

Annual input data for NEAC PFA run-reconstruction & NCL models for UK(ENGLAND AND WALES). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Estimated	Declared	Declared	Declared	Estimated	Estimated		Estimated		Estimated	Uncertainty	Estimated	Uncertainty	Estimated	Estimated	Estimated	Estimated
rear	total catch	proportion		catch in				Uncertainty		Uncertainty	exploitation		exploitation	in			proportion	
	total catell	1SW (total)		NE coastal			unreported	•	unreported	in %	•	exploitation	•	exploitation		Scottish	Scottish	Scottish
		2011 (1014.)	fishery	fishery -		fishery)	•	unreported	•		1SW salmon	rate (%) -	MSW	•	catch in NE		fish in NE	
			total	•	•	,,	1SW	catch of	MSW	catch of		1SW salmon	salmon	MSW	fishery	fishery		fishery (T/J
1971	109 861	0.55	60 353		,	0.55	38.3	9.6	38.3	9.6	57.3	10.0	42.5	10.0	32.3	0.95		/ ( / -
1972	108 074	0.42	51 681			0.42	39.0	9.7	39.0	9.7	51.3	10.0	37.8	10.0	32.3	0.95		
1973	114 786	0.53	62 842			0.53	38.4	9.6	38.4	9.6	50.6	10.0	37.3	10.0	32.3	0.95		
1974	104 325	0.65	52 756			0.65	39.3	9.8	39.3	9.8	50.2	10.0	37.0	10.0	32.3	0.95		
1975	113 062	0.59	53 451			0.59	38.5	9.6	38.5	9.6	49.8	10.0	36.7	10.0	32.3	0.95		
1976	54 294	0.64	15 701			0.64	36.8	9.2	36.8	9.2	50.3	10.0	37.1	10.0	32.3	0.94		
1977	94 282	0.62	52 888			0.62	39.0	9.8	39.0	9.8	50.4	10.0	37.2	10.0	32.3	0.93		
1978	93 125	0.69	51 630			0.69	38.4	9.6	38.4	9.6	49.1	10.0	36.2	10.0	32.3	0.92		
1979	75 386	0.81	43 464			0.81	38.6	9.6	38.6	9.6	47.7	10.0	35.2	10.0	32.3	0.91		
1980	90 218	0.55	45 780			0.55	39.1	9.8	39.1	9.8	47.8	10.0	35.2	10.0	32.3	0.90		
1981	121 039	0.48	69 113			0.48	38.3	9.6	38.3	9.6	47.4	10.0	34.9	10.0	32.3	0.89		
1982	80 289	0.67	50 167			0.67	38.3	9.6	38.3	9.6	47.3	10.0	34.8	10.0	32.3	0.88		
1983	116 995	0.72	77 277			0.72	37.1	9.3	37.1	9.3	47.1	10.0	34.7	10.0	32.3	0.87		
1984	94 271	0.74	59 295			0.74	36.5	9.1	36.5	9.1	47.4	10.0	34.8	10.0	32.3	0.86		
1985	95 531	0.66	57 356			0.66	38.9	9.7	38.9	9.7	47.5	10.0	34.9	10.0	32.3	0.85		
1986	110 794	0.62	63 425			0.62	38.0	9.5	38.0	9.5	46.9	10.0	34.3	10.0	32.3	0.84		
1987	83 439	0.68	36 143			0.68	38.2	9.5	38.2	9.5	46.1	10.0	33.7	10.0	32.3	0.83		
1988	110 163	0.69		47 465	3 384	0.69	39.7	9.9	39.7	9.9	45.5	10.0	33.5	10.0	32.3		0.82	0.50
1989	83 668	0.65		36 236	5 217	0.65	36.9	9.2	36.9	9.2	45.3	10.0	33.3	10.0	32.3		0.81	0.50
1990	86 676	0.52		48 219	3 311	0.52	36.7	9.2	36.7	9.2	45.3	10.0	33.2	10.0	31.3		0.80	0.50
1991	51 649	0.71		22 463	2 966	0.71	37.3	9.3	37.3	9.3	44.0	10.0	32.3	10.0	29.7		0.79	0.50
1992	44 586	0.77		17 574	2 570	0.77	39.8	10.0	39.8	10.0	43.5	10.0	31.8	10.0	28.0		0.78	0.50
1993	69 177	0.81		39 224	2 576	0.81	38.0	9.5	38.0	9.5	40.6	10.0	29.5	10.0	26.3		0.77	0.50
1994	88 121	0.77		41 298	5 256	0.77	23.9	6.0	23.9	6.0	40.5	10.0	29.5	10.0	24.4		0.76	0.50
1995	80 478	0.72		48 005	5 205	0.72	22.3	5.6	22.3	5.6	37.6	10.0	27.1	10.0	22.5		0.75	0.50
1996	46 696	0.65		15 172	3 409	0.65	20.6	5.1	20.6	5.1	35.8	10.0	25.8	10.0	20.6		0.75	0.50
1997	41 374	0.73		19 241	2 681	0.73	18.8	4.7	18.8	4.7	33.4	10.0	23.9	10.0	18.5		0.75	0.50
1998	36 917	0.82		17 328	937	0.82	18.9	4.7	18.9	4.7	31.4	10.0	22.4	10.0	18.5		0.75	0.50
1999	41 094	0.68		24 812	2 021	0.68	17.4	4.4	17.4	4.4	29.5	10.0	17.9	9.0	17.1		0.75	0.50
2000	60 953	0.79		40 059	3 295	0.79	14.9	3.7	14.9	3.7	29.7	10.0	15.0	7.5	13.1		0.75	0.50
2001	51 307	0.75		32 374	3 741	0.75	14.8	3.7	14.8	3.7	27.9	10.0	14.3	7.1	13.1		0.75	0.50
2002	45 669	0.76		27 685	3 295	0.76	15.3	3.8	15.3	3.8	27.8	10.0	14.1	7.0	13.9		0.75	0.50
2003	22 206	0.66		5 511	4 924	0.66	17.4	4.4	17.4	4.4	21.4	10.0	10.7	5.3	17.1		0.75	0.50
2004	30 559	0.81		5 921	5 096	0.81	17.7	4.4	17.7	4.4	22.1	10.0	10.6	5.3	17.1		0.75	0.50
2005	26 162	0.76		5 607	3 380	0.76	17.6	4.4	17.6	4.4	21.8	10.0	10.6	5.3	17.1		0.75	0.50

2006	22 056	0.78	4 040	3 526	0.78	17.6	4.4	17.6	4.4	19.5	9.8	9.1	4.6	17.1	0.75	0.50
2007	19 914	0.78	4 894	2 197	0.78	17.7	4.4	17.7	4.4	17.9	9.0	8.4	4.2	17.1	0.75	0.50
2008	19 036	0.76	3 649	2 592	0.76	17.8	4.4	17.8	4.4	17.6	8.8	8.2	4.1	17.1	0.75	0.50
2009	13 910	0.72	2 590	2 805	0.72	11.4	2.9	11.4	2.9	17.4	8.7	8.2	4.1	7.4	0.75	0.50
2010	32 695	0.78	12 214	7 768	0.78	10.8	2.7	10.8	2.7	17.5	8.8	8.0	4.0	7.4	0.75	0.50
2011	34 575	0.57	14 915	9 233	0.57	10.1	2.5	10.1	2.5	20.8	10.0	10.2	5.1	7.4	0.64	0.37
2012	14 926	0.50	3 571	3 705	0.50	11.2	2.8	11.2	2.8	16.8	8.4	8.0	4.0	7.4	0.62	0.37
2013	22 608	0.58	7 964	8 679	0.58	9.5	2.4	9.5	2.4	17.4	8.7	8.5	4.3	7.4	0.63	0.37
2014	14 219	0.54	6 974	3 826	0.54	9.3	2.3	9.3	2.3	15.8	7.9	8.0	4.0	7.4	0.64	0.37
2015	19 262	0.47	9 233	6 657	0.47	12.9	3.2	12.9	3.2	15.2	7.6	7.7	3.9	7.4	0.64	0.37
2016	22 494	0.42	10 811	7 956	0.42	12.0	3.0	12.0	3.0	14.8	7.4	7.5	3.7	7.4	0.63	0.37
2017	12 164	0.40	5 095	3 975	0.40	13.8	3.4	13.8	3.4	11.7	5.8	5.8	2.9	7.4	0.63	0.38

### UK (Northern Ireland)-Foyle Fisheries Area

Annual input data for NEAC PFA run-reconstruction & NCL models for UK(NORTHERN IRELAND) (FOYLE). (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Year	Declared	Declared	Declared rod	Declared rod	Estimated %		Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in
	net catch	net catch	catch 1SW	catch MSW	unreported	% unreported	unreported	% unreported	exploitation	exploitation	exploitation	exploitation
	1SW salmon	MSW	salmon	salmon	catch of 1SW	catch of 1SW	catch of MSW	catch of MSW	rate (%) - 1SW	rate (%) - 1SW	rate (%) - MSW	
		salmon			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon
1971	78 037	5 874			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1972	64 663	4 867			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1973	57 469	4 326			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1974	72 587	5 464			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1975	51 061	3 843			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1976	36 206	2 725			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1977	36 510	2 748			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1978	44 557	3 354			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1979	34 413	2 590			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1980	45 777	3 446			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1981	32 346	2 435			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1982	55 946	4 211			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1983	77 424	5 828			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1984	27 465	2 067			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1985	37 685	2 836			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1986	43 109	3 245			21.5	11.5	21.5	11.5	80.0	5.0	50.0	5.0
1987	17 189	1 294			21.5	11.5	21.5	11.5	69.0	7.0	46.0	5.0
1988	43 974	3 310			21.5	11.5	21.5	11.5	64.5	6.5	36.0	4.0
1989	60 288	4 538			23.5	13.5	23.5	13.5	89.0	9.0	60.0	6.0
1990	39 875	3 001			13.5	3.5	13.5	3.5	62.0	6.0	38.0	4.0
1991	21 709	1 634			13.5	3.5	13.5	3.5	64.5	6.5	43.0	4.0
1992	39 299	2 958			16.5	6.5	16.5	6.5	56.0	6.0	33.0	3.0
1993	35 366	2 662			13.5	3.5	13.5	3.5	41.0	4.0	12.0	1.0
1994	36 144	2 720			19.0	9.0	19.0	9.0	70.0	7.0	40.0	4.0
1995	33 398	2 514			13.5	3.5	13.5	3.5	67.0	7.0	42.0	4.0
1996	28 406	2 138			15.0	5.0	15.0	5.0	57.0	10.0	34.0	10.0
1997	40 886	3 077			10.0	5.0	10.0	5.0	60.0	10.0	34.0	10.0
1998 1999	37 154 21 660	2 797 1 630			10.0 10.0	5.0 5.0	10.0 10.0	5.0 5.0	25.0 63.0	5.0 5.0	22.5 32.5	7.5 7.5
							10.0	5.0	58.0			
2000 2001	30 385 21 368	2 287 1 608			10.0 5.0	5.0 5.0	5.0		50.0	5.0 5.0	32.5 30.0	7.5 5.0
2001	37 914	2 854	9 163	690	2.5	2.5	2,5	2.5	15.0	3.0	15.0	3.0
2002	30 441	2 291	4 576	344	0.5	0.5	0.5	0.5	15.0	3.0	15.0	3.0
2003	20 730	1 560	4 570	344	0.5	0.5	0.5	0.5	15.0	3.0	15.0	3.0
2004	23 746	1 787	7 079	533	0.5	0.5	0.5	0.5	15.0	3.0	15.0	3.0
2005	11 324	852	4 886	368	0.5	0.5	0.5	0.5	15.0	3.0	15.0	3.0
2007	5 050	322	9 530	608	0.5	0.5	0.5	0.5	15.0	3.0	15.0	3.0
2008	3 880	292	4 755	304	0.5	0.5	0.5	0.5	15.0	3.0	15.0	3.0
2009	1 743	194	3 640	405	0.5	0.5	0.5	0.5	15.0	3.0	15.0	3.0
2010	0	0	3 488	388	0.5	0.5	0.5	0.5	15.0	3.0	15.0	3.0
2010	0	0	2 276	759	1.0	1.0	1.0	1.0	15.0	5.0	15.0	5.0
2011	0	0	4 781	1594	1.0	1.0	1.0	1.0	10.0	7.5	10.0	7.5
2012	0	0	5 030	498	1.0	1.0	1.0	1.0	10.0	7.5	10.0	7.5
2013	0	0	2 029	225	1.0	1.0	1.0	1.0	10.0	7.5	10.0	7.5
2014	0	0	1 998	250	1.0	1.0	1.0	1.0	10.0	7.5	10.0	7.5
2015	0	0	3 192	355	1.0	1.0	1.0	1.0	10.0	7.5	10.0	7.5
2010	0	0	2 101	233	1.0	1.0	1.0	1.0	10.0	7.5	10.0	7.5

#### UK (Northern Ireland)-DAERA area

Annual input data for NEAC PFA run-reconstruction & NCL models for UK(NORTHERN IRELAND) (LOCHS AGENCY AREA) (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation)

Yea	Declared	Declared	Declared rod	Declared rod	Estimated %	Uncertainty in	Estimated %	Uncertainty in	Estimated	Uncertainty in	Estimated	Uncertainty in C	ount River	Estimated	Count	Estimated	Count	Uncertainty	Count	Uncertainty
	net catch	net catch	catch 1SW	catch MSW		% unreported	unreported	•	exploitation	exploitation	exploitation	exploitation	Bush			proportion		in upscaling		
	1SW salmon	MSW	salmon	salmon							rate (%) - MSW			(%) 1SW in		(%) 1SW in		factor (%)	factor	factor (%)
		salmon			salmon	salmon	salmon	salmon	salmon	salmon	salmon	salmon		Bush		Bann	(%) 1SW	1SW	(%) MSW	MSW
197		2 673			21.5	11.5	21.5	11.5	80	5	50	5	C				-			
197		2 601			21.5	11.5	21.5	11.5	80	5	50	5	(				-			
197		2 200			21.5	11.5	21.5	11.5	80	5	50	5	(							
197		1 679			21.5	11.5	21.5	11.5	80	5	50	5			, ,	(	-			
197		2 010			21.5	11.5	21.5	11.5	80	5	50	5	9							
197		1 346			21.5	11.5	21.5	11.5	80	5	50	5					-			
197 197		1 263			21.5	11.5	21.5	11.5	80 80	5 5	50 50	5		) (	, ,	7				
		1871			21.5	11.5	21.5	11.5		5	50	5		,	,	7				
197 198		1 078 1 202			21.5 21.5	11.5 11.5	21.5 21.5	11.5 11.5	80 80	5	50	3		) (						
					21.5	11.5	21.5	11.5	80	5	50			,		(				
198 198		1 204 1 059			21.5	11.5	21.5	11.5	80	5	50	3	• (	,		(				
198		1 569			21.5	11.5	21.5	11.5	80	5	50	5	(				-			
198		836			21.5	11.5	21.5	11.5	80	J	50	5	(							
198		931			21.5	11.5	21.5	11.5	80	5	50	5	(				-			
198		991			21.5	11.5	21.5	11.5	80	5	50	5	(			7	-			
198		695			21.5	11.5	21.5	11.5	00	4	10	3	2 530					2.0	9.5	2.0
198		1 078			21.5	11.5	21.5	11.5		`			2 832							
198		1 135			23.5	13.5	23.5	13.5					1 029							
199		715			13.5	3.5	13.5	3.5					1 850							
199		526			13.5	3.5	13.5	3.5					2 341							
199		703			16.5	6.5	16.5	6.5					2 546			(				
199		595			13.5	3.5	13.5	3.5					3 235							
199	11 206	843			19.0	9.0	19.0	9.0					2 010		3 0	(	10.0	2.0	9.5	
199	11 637	876			13.5	3.5	13.5	3.5					1 521	92	2 0	(	10.0	2.0	9.5	2.0
199		781			15.0	5.0	15.0	5.0					1 097			(				
199	10 479	789			10.0	5.0	10.0	5.0					1 677	85	6 541	85	67.0	5.0	61.0	5.0
199	9 375	706			10.0	5.0	10.0	5.0					2 995	95	5 11 462	95	67.0	5.0	61.0	5.0
199	9 011	678			10.0	5.0	10.0	5.0					977	90	3 599	90	67.0	5.0	61.0	5.0
200	10 598	798			10.0	5.0	10.0	5.0					950	91	1 5 979	91	1 67.0	5.0	61.0	5.0
200	8 104	610			5.0	5.0	5.0	5.0					913	97	7 5 771	97	7 67.0	5.0	61.0	5.0
200	3 3 1 5	249	2 218	167	2.5	2.5	2.5	2.5					835	95	5 5 037	95	67.0	5.0	61.0	5.0
200	2 236	168	1 884	141	2.5	2.5	2.5	2.5					723	96	5 4 147	96	67.0	5.0	61.0	5.0
200	2 411	181	3 053	230	0.5	0.5	0.5	0.5					878	92	9 050	92	2 67.0	5.0	61.0	5.0
200	3 012	227	1 791	135	0.5	0.5	0.5	0.5					1 151	. 91	L 6 609	91	1 67.0	5.0	61.0	5.0
200	2 288	172	1 289	97	0.5	0.5	0.5	0.5					1 074	1 87	7 410	87	7 67.0	5.0	61.0	5.0
200	2 533	162	2 427	155	0.5	0.5	0.5	0.5					2 584	1 94	7 008	94	1 67.0	5.0	61.0	5.0
200		116	2 444	156	0.5	0.5	0.5	0.5					1 712			(	10.0			
200		154	1 457	162	0.5	0.5	0.5	0.5					726							
201		191	1 327	147	0.5	0.5	0.5	0.5					1 045							
201		285	1 132	378	1.0	1.0	1.0	1.0					649							
201		5	263	87	1.0	1.0	1.0	1.0					926							
201		1	46	5	1.0	1.0	1.0	1.0					1 644							
201		0	143	40	2.5	2.5	2.5	2.5					963							
201		0	20	6	2.5	2.5	2.5	2.5					1 005							
201		0	124	279	2.5	2.5	2.5	2.5					2 166							
201	0	0	104	208	2.5	2.5	2.5	2.5					912	! 88	8 433	81	1 67.0	5.0	61.0	5.0

### UK (Scotland)-East

catch 1S' salmo  1971	ch 1SW	eleased rod	catch 1SW													
1971				correction	correction	proportion	estimated	retained rod	released rod		correctio	correction	unreported catch of		eggs per	
1971 4 58 1972 4 67 1973 5 27 1974 5 97 1975 4 71 1976 5 28 1977 6 64 1978 7 91 1979 10 76 6 91 1980 7 33 1981 8 40 1982 12 41 1983 96 7 13 1984 10 55 1985 12 42 1986 11 51 1987 13 71 1988 19 26 1989 21 25 1990 13 94 1991 12 54 1991 12 54 1992 21 54 1993 22 88 1994 19 41 1995 18 65 1993 12 86 1994 19 41 1995 18 65 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 6 82 2006 15 64 2007 13 69 2009 8 30 2009 8 30		catch 1SW	salmon	factor 1SW	factor 1SW	unreported	proportion	catch MSW	catch MSW	salmon	n factor	factor MSW	MSW salmon	unreported catch of	female	femal
1972	aiiiioii	salmon		salmon	salmon	catch of 1SW	unreported	salmon	salmon		MSW	salmon		MSW salmon	1SW	MS۱
1972						salmon	catch of 1SW salmon				salmon					
1972	4 581	0	212 292	0.012763	0.000805	0.10	0.05	34 192	0	101 338	0.152792	0.011450	0.10	0.05	4 954	7 28
1973 5 27 1974 5 97 1975 4 71 1975 4 71 1976 5 28 1977 6 64 1978 7 91 1979 10 76 1980 7 33 1981 8 40 1982 12 41 1983 96 7 1984 10 55 1985 12 42 1986 11 51 1987 13 71 1988 19 26 1989 21 25 1990 13 94 1991 12 54 1993 22 88 1994 19 41 1995 18 65 1997 14 44 1998 22 79 1999 10 11 1996 12 54 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 190 2002 11 31 2003 6 82 2005 13 48 2006 15 64 2007 13 69 2008 10 05 2009 8 30 2010 12 91	4 672	0	215 434	0.011456	0.000832	0.10	0.05	45 211	0		0.132752	0.010715	0.10	0.05	5 092	7.5
1974 5 97 1975 4 71 1975 4 71 1976 5 28 1977 6 64 1978 7 91 1979 10 76 1980 7 33 1981 8 40 1982 12 41 1983 9 67 1986 11 51 1987 13 71 1988 19 26 1990 13 94 1991 12 54 1992 21 54 1993 22 88 1994 19 41 1993 22 88 1994 19 41 1995 18 65 1996 16 86 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 6 82 2004 14 54 2007 13 69 2008 10 05 2009 8 30 2009 8 80 2009 18 2009 8 80	5 277	0	254 496	0.010567	0.000786	0.10	0.05	49 569	0		0.163465	0.011905	0.10	0.05	5 193	77
1975 471 1976 528 1977 664 1978 791 1979 1076 1980 733 1981 840 1982 12 41 1983 967 1984 10 55 1984 10 55 1985 11 51 1987 13 71 1988 19 26 1990 13 94 1991 12 54 1992 21 54 1993 22 88 1994 19 41 1995 18 65 1996 16 86 1996 16 86 1997 14 44 1998 22 79 1999 10 11 2000 14 14 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 682 2004 14 54 2007 13 69 2008 10 05 2009 8 30 2009 8 80	5 971	0	239 453	0.012882	0.000976	0.10	0.05	41 764	0		0.184708	0.012303	0.10	0.05	5 241	7 86
1977 6 64 1978 7 91 1979 10.76 1980 7 33 1981 8 40 1982 12 41 1983 967 1985 12 42 1986 11 51 1987 13 71 1988 19 26 1989 21 25 1990 13 94 1991 12 54 1993 22 88 1994 19 41 1995 18 65 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 6 82 2004 14 54 2005 13 48 2006 15 64 2007 13 69 2008 10 05 2008 10 05 2009 8 30 2010 12 91	4 718	0	177 222	0.011925	0.001047	0.10	0.05	54 153	0		0.177313	0.014648	0.10	0.05	5 227	7 94
1977 6 64 1978 7 91 1979 10.76 1980 7 33 1981 8 40 1982 12 41 1983 967 1985 12 42 1986 11 51 1987 13 71 1988 19 26 1989 21 25 1990 13 94 1991 12 54 1993 22 88 1994 19 41 1995 18 65 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 6 82 2004 14 54 2005 13 48 2006 15 64 2007 13 69 2008 10 05 2008 10 05 2009 8 30 2010 12 91	5 287	0	144 782	0.017207	0.001647	0.10	0.05	33 770	0		0.192407	0.017829	0.10	0.05	5 153	7 9
1979 10.76 1980 7.33 1981 8.40 1982 12.41 1983 9.67 1984 10.55 1985 12.42 1986 11.51 1987 13.71 1988 19.26 1990 13.94 1991 12.54 1992 21.54 1993 22.88 1994 19.91 1995 18.65 1996 16.86 1996 16.86 1996 16.80 1997 14.44 1998 22.79 1999 10.11 2000 14.14 1998 22.79 1999 10.11 2000 14.49 2001 14.90 2002 11.31 2003 6.82 2004 14.54 2007 13.69 2006 15.64 2007 13.69 2009 8.30 2009 8.80	6 648	0	147 658	0.018069	0.001573	0.10	0.05	49 419	0	69 226	0.209492	0.017423	0.10	0.05	5 033	7 89
1980 7 33 1981 8 40 1982 12 41 1983 9 67 1984 10 55 1985 12 42 1986 11 51 1987 13 71 1988 19 26 1989 21 25 1990 13 34 1991 12 54 1992 21 54 1993 22 88 1994 19 41 1995 18 65 1997 14 44 1998 22 79 1999 10 11 2003 62 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2009 8 30 2010 12 91	7 913	0	150 946	0.021422	0.001701	0.10	0.05	59 080	0		0.174402	0.015128	0.10	0.05	4 886	7 80
1981 8 40 1982 12 41 1983 967 1984 10 55 1985 12 42 1986 11 51 1987 13 71 1988 19 26 1989 21 25 1990 13 94 1991 12 54 1993 22 88 1994 19 41 1995 18 65 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 6 22 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2007 13 69 2008 10 05 2009 8 30 2010 12 91	10 760	0	150 036	0.027871	0.002244	0.10	0.05	58 124	0		0.215980	0.018674	0.10	0.05	4 732	7 70
1982 12 41 1983 9 67 1984 10 55 1985 12 42 1986 11 51 1987 13 71 1988 19 26 1990 13 94 1991 12 54 1992 21 54 1993 22 88 1994 19 41 1995 18 65 1996 16 86 1996 16 86 2001 14 14 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 6 82 2004 14 54 2007 13 69 2008 10 05 2009 8 30 2010 12 91	7 336	0	94 329	0.030202	0.002178	0.10	0.05	52 184	0		0.151975	0.012201	0.10	0.05	4 593	7 62
1983 9 67 1984 10 55 1985 12 42 1986 11 51 1987 13 71 1988 19 26 1990 13 94 1991 12 54 1992 21 54 1992 22 15 1993 12 28 1994 19 41 1995 18 65 1996 16 86 1996 16 86 2001 14 40 2002 11 31 2003 62 2004 14 54 2007 13 69 2008 10 05 2009 8 30 2010 12 91	8 409	0	121 281	0.027035	0.002212	0.10	0.05	42 896	0	113 787	0.153269	0.010262	0.10	0.05	4 484	7 56
1984 10 55 1985 12 42 1986 11 51 1987 13 71 1988 19 26 1989 21 25 1990 13 94 1991 12 54 1992 21 54 1993 22 88 1994 19 41 1995 18 65 1996 16 86 1997 14 44 1998 22 79 1001 11 2000 14 14 2001 14 90 2002 11 31 2003 68 22 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2007 13 69 2007 13 69 2009 8 30 2010 12 91	12 417	0	162 957	0.034158	0.002423	0.10	0.05	40 398	0	72 800	0,203141	0.015532	0.10	0.05	4 413	7 54
1985 12 42 1986 11 51 1987 13 71 1988 19 26 1989 21 25 1990 13 94 1991 12 54 1993 22 88 1994 19 41 1995 18 65 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 62 2004 14 54 2005 13 48 2006 15 64 2007 13 69 2007 13 69 2007 13 69 2008 8 30 2010 12 91	9 670	0	161 173	0.024506	0.002121	0.10	0.05	43 671	0	82 433	0.203469	0.015812	0.10	0.05	4 383	7 5
1986 11 51 1987 13 71 1988 19 26 1989 21 25 1990 13 94 1991 12 54 1992 21 54 1995 18 65 1995 16 86 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 68 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2008 8 30 2010 12 91	10 557	0	165 118	0.028029	0.002596	0.10	0.05	36 321	0	54 508	0.192885	0.018928	0.10	0.05	4 393	7 60
1987 13 71 1988 19 26 1989 21 25 1990 13 94 1991 12 54 1992 21 54 1993 22 88 1994 19 41 1995 18 65 1996 16 86 1996 16 86 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 6 22 2004 14 54 2007 13 69 2008 10 05 2009 8 30 2010 12 91	12 427	0	120 692	0.035779	0.002763	0.10	0.05	47 258	0	47 786	0.238414	0.023174	0.10	0.05	4 433	7 6
1988 19 26 1989 21 25 1990 13 94 1991 12 54 1992 21 54 1993 22 88 1994 19 41 1995 18 65 1996 16 86 1997 14 44 1998 22 79 100 11 2000 14 14 2001 14 90 2002 11 31 2003 68 22 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2007 13 69 2008 10 05 2008 83 2010 12 91	11 519	0	168 773	0.027069	0.002100	0.10	0.05	48 519	0	80 135	0.200841	0.015981	0.10	0.05	4 492	7.7
1989 21 25 1990 13 94 1991 12 54 1992 21 54 1993 22 88 1994 19 41 1995 18 65 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 62 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2007 13 69 2008 10 05 2009 8 30 2010 12 91	13 710	0	125 549	0.037201	0.002774	0.10	0.05	44 326	0	44 205	0.251613	0.021974	0.10	0.05	4 554	77
1990 13 94 1991 12 54 1992 21 54 1993 22 88 1994 19 41 1995 18 65 1996 16 86 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 6 82 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2008 10 05 2009 8 30 2010 12 91	19 262	0	99 358	0.044678	0.003894	0.10	0.05	53 778	0	37 389	0.316713	0.028043	0.10	0.05	4 606	7 7
1991 12 54 1992 21 54 1993 22 88 1994 19 41 1995 18 65 1996 16 86 1996 16 97 1999 10 11 2000 14 14 2001 14 99 2002 11 31 2003 6 82 2004 14 54 2007 13 69 2008 10 05 2008 10 05 2009 8 30 2010 12 91	21 251	0	121 812	0.040752	0.003962	0.10	0.05	46 689	0	38 710	0.272810	0.025769	0.10	0.05	4 633	7.7
1992 21 54 1993 22 88 1994 19 41 1995 18 65 1996 16 86 1997 14 44 1998 22 79 1001 14 2001 14 90 2002 11 31 2003 68 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2007 13 69 2008 10 005 2009 8 30 2010 12 91	13 946	0	49 406	0.042769	0.004104	0.10	0.05	42 634	0	31 340	0.240908	0.023612	0.10	0.05	4 629	76
1993 22 88 1994 19 41 1995 18 65 1996 16 86 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 62 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2007 13 69 2008 10 05 2009 8 30 2010 12 91	12 544	0	41 317	0.045369	0.005005	0.10	0.05	37 497	0	16 196	0.240940	0.024636	0.10	0.05	4 590	7 60
1994 19 41 1995 18 65 1996 16 86 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 6 82 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2008 10 05 2009 8 30 2010 12 91	21 544	0	58 362	0.056558	0.004735	0.10	0.05	45 154	0	22 825	0.322544	0.026275	0.10	0.05	4 520	75
1995 18 65 1996 16 86 1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 68 22 2004 16 54 2005 13 43 2006 15 64 2007 13 69 2008 10 05 2009 8 30 2010 12 91	22 888	0	50 525	0.054677	0.006083	0.10	0.05	43 860	0	16 638	0.302020	0.027928	0.10	0.05	4 427	7 43
1996	19 418	1 295	61 011	0.046119	0.004753	0.10	0.05	45 550	4 634	27 208	0.261093	0.025712	0.10	0.05	4 322	7 34
1997 14 44 1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 68 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2007 13 69 2007 13 69 2008 10 05 2009 8 30 2010 12 91	18 650	2 217	54 370	0.047264	0.005164	0.10	0.05	45 935	8 267	23 120	0.257429	0.023982	0.10	0.05	4 215	7 26
1998 22 79 1999 10 11 2000 14 14 2001 14 90 2002 11 31 2003 682 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2009 8 30 2010 12 91	16 869	1 716	39 758	0.050523	0.005363	0.10	0.05	34 573	7 402	15 792	0.230297	0.022580	0.10	0.05	4 114	7 19
1999 10 11 2000 14 14 2001 14 90 2001 13 90 2002 11 31 2003 6 82 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2008 10 05 2009 8 30 2010 12 91	14 445	2 228	23 003	0.055266	0.005748	0.10	0.05	28 128	7 400	6 722	0.276574	0.026546	0.10	0.05	4 026	7 11
2000 14 14 2001 14 90 2002 11 31 2003 6 82 2004 14 54 2005 13 43 2006 15 64 2007 13 69 2008 10 05 2009 8 30 2010 12 91	22 797	4 337	22 155	0.079107	0.007336	0.10	0.05	27 439	7 721	4 792	0.324222	0.032765	0.10	0.05	3 953	7 02
2001     14 90       2002     11 31       2003     6 82       2004     14 54       2005     13 43       2006     15 64       2007     13 69       2008     10 05       2009     8 30       2010     12 91	10 113	3 020	10 794	0.059927	0.005846	0.10	0.05	22 140	10 185	4 871	0.263481	0.026976	0.10	0.05	3 894	6 92
2002     11 31       2003     6 82       2004     14 54       2005     13 43       2006     15 64       2007     13 69       2008     10 05       2009     8 30       2010     12 91	14 143	5 967	22 728	0.059465	0.006271	0.10	0.05	22 630	12 306	8 650	0.279134	0.029195	0.10	0.05	3 848	6 82
2003     6 82       2004     14 54       2005     13 43       2006     15 64       2007     13 69       2008     10 05       2009     8 30       2010     12 91	14 900	7 235	21 746	0.060723	0.005904	0.10	0.05	22 571	16 689	7 899	0.228555	0.025205	0.10	0.05	3 810	67
2004     14 54       2005     13 43       2006     15 64       2007     13 69       2008     10 05       2009     8 30       2010     12 91		6 520	15 301	0.068607	0.006268	0.10	0.05	16 141	13 830	5 599	0.258985	0.027854	0.10	0.05	3 776	6 63
2005     13 43       2006     15 64       2007     13 69       2008     10 05       2009     8 30       2010     12 91	6 823	7 651	19 048	0.050968	0.005348	0.10	0.05	12 827	18 255	11 443	0.216467	0.025033	0.10	0.05	3 744	6 53
2006     15 64       2007     13 69       2008     10 05       2009     8 30       2010     12 91		12 722	17 124	0.072754	0.006629	0.10	0.05	23 284	27 819	7 489	0.256732	0.027153	0.10	0.05	3 709	6 47
2007 13 69 2008 10 05 2009 8 30 2010 12 91		12 633	18 160	0.069414	0.007054	0.10	0.05	17 424	26 039	6 252	0.228581	0.022968	0.10	0.05	3 673	6.4
2008 10 05 2009 8 30 2010 12 91		16 344	15 090	0.092951	0.009012	0.10	0.05	16 298	25 854	6 656	0.179359	0.017929	0.10	0.05	3 636	6 4
2009 8 30 2010 12 91		18 807	12 316	0.097428	0.009752	0.10	0.05	14 907	26 975	4 537	0.217876	0.022229	0.10	0.05	3 600	6 4
2010 12 91		14 458	8 536	0.090754	0.008312	0.10	0.05	15 557	31 238	5 200	0.180614	0.018003	0.10	0.05	3 568	65
	8 302	13 334	6 561	0.101537	0.009086	0.10	0.05	10 641	28 757	4 401	0.181367	0.017636	0.10	0.05	3 544	65
2011 6 21		28 151	15 250	0.108500	0.009661	0.10	0.05	13 464	41 010	9 397	0.193488	0.018107	0.10	0.05	3 531	6 5
	6 214	11 762	6 281	0.093604	0.008623	0.10	0.05	12 210	43 824	12 008	0.163768	0.017382	0.10	0.05	3 529	66
	7 347	16 755	8 785	0.097039	0.009101	0.10	0.05	10 124	37 475	6 100	0.175273	0.016478	0.10	0.05	3 541	65
		12 858	14 126	0.082192	0.007886	0.10	0.05	6 307	34 519	8 594	0.161116	0.016182	0.10	0.05	3 564	65
	4 274		8 407	0.096629	0.008899	0.10	0.05	3 810	24 008	7 988	0.167562	0.016377	0.10	0.05	3 597	64
	2 517	8 077									0.149389	0.016062	0.10			6 38
2016 2 58 2017 1 93	2 517 3 858	14 642 12 807	8 365 1 441	0.098923 0.085701	0.010004 0.008156	0.10 0.10	0.05 0.05	3 589 2 637	26 394 29 705	4 203 1 370	0.149589	0.010002	0.10	0.05 0.05	3 636 3 679	6 2

#### UK (Scotland)-West

2017

3279

33 0.0699304

0.0071636

0.10

0.05

3 797

79 0.1310814 0.016983671

0.10

0.05

3 722

6 194

Annual input data for NEAC PFA run-reconstruction & NCL models for SCOTLAND (WEST) (Uncertainty values define normal distribution around estimates used in Monte Carlo simulation) Declared Declared Declared Estimated Declared Declared Mean Uncertainty Estimated Uncertainty in Mean Mean Uncertainty in Uncertainty in Declared net Mean retained rod released rod net catch correction correction proportion estimated retained rod released rod catch MSW correction in correction proportion eggs per eggs per catch 1SW catch 1SW 1SW factor 1SW factor 1SW unreported proportion catch MSW catch MSW salmon factor MSW factor MSW female female unreported proportion catch of 1SW unreported MSW salmon salmor salmon salmor catch of unreported 1SW salmon catch of 1SW MSW salmon catch of MSW salmon salmon 1971 3 497 0 41790 0.0236804 0.0022187 0.10 0.05 7 255 18 816 0.14576203 0.016216218 0.10 0.05 4 954 7 285 2 078 0.0173803 26 464 0.12816127 0.010252915 7 517 1972 0 29280 0.0014502 0.10 0.05 7 684 0.10 0.05 5 092 24 129 0.17488261 0.016924578 2 495 30822 0.0169505 0.0017887 8 965 5 193 7 718 1973 0 0.10 0.05 0.10 0.05 20 818 0.18121809 0.018668083 1974 3 605 0 40387 0.0225981 0.0023418 0.10 0.05 8 551 0.10 0.05 5 241 7 865 0.0211948 0.0017220 6 862 20 198 0.15207771 0.013738818 0.10 0.05 5 227 7 943 1975 2 5 1 0 37707 0.10 0.05 15 243 0.19003155 0.01681585 1976 2 5 1 8 0 35736 0.0211604 0.0015372 0.10 0.05 7 049 0.10 0.05 5 153 7 949 13 929 0.16254177 0.017033098 Ω 6 357 0.10 0.05 5.033 7 897 1977 2 130 37609 0.0187729 0.0014665 0.10 0.05 1978 3 357 Ω 41985 0.0255894 0.0021948 0.10 0.05 7 007 16 078 0.15692116 0.015246973 0.10 0.05 4 886 7 808 1979 4 484 0 21800 0.0392930 0.0035625 0.10 0.05 7 797 8 126 0.18830465 0.022284963 0.10 0.05 4 732 7 708 729 0.13502385 0.020597835 1980 3 831 15823 0.0398125 0.0043158 0.05 7 153 0.10 0.05 4 593 7 621 0 0.10 803 0.11243399 0.01798228 1981 3 863 0 16779 0.0310103 0.0043758 0.10 0.05 8 093 0.10 0.05 4 484 7 563 488 0.19555802 0.02485174 4 413 1982 4 4 2 2 Ω 27930 0.0333143 0.0033048 0.10 0.05 7 517 0.10 0.05 7 542 4 439 0.0327860 11 488 0.22681907 0.02035421 0.05 4 383 1983 0 33967 0.0026245 0.10 0.05 8 290 0.10 7 557 9 552 0.13190637 0.026173908 1984 4 968 32294 0.0258663 0.0052748 0.10 0.05 6 769 0.10 0.05 4 393 7 600 1985 5 222 0 19426 0.0391525 0.0034227 0.10 0.05 11 183 8 338 0.26603804 0.0320413 0.10 0.05 4 433 7 656 Ω 18283 8 784 0 1944814 0 023420486 0.05 4 492 1986 4 200 0.0259341 0.0025706 0.10 0.05 0.10 7 709 1987 4 350 Ω 21038 0.0337496 0.0028095 0.10 0.05 6706 0.28514277 0.032475402 0.10 0.05 4 554 7 744 1988 8 547 0 21765 0.0391303 0.0042226 0.10 0.05 14 901 6 071 0.42169133 0.056563204 0.10 0.05 4 606 7 752 1989 8 4 1 8 0 23370 0.0370481 0.0039132 0.10 0.05 11 649 6 825 0.33589737 0.035589483 0.10 0.05 4 633 7 728 1990 5 420 O 12309 0.0378462 0.0036981 0.10 9 646 4 296 0.29852251 0.036808674 0.10 0.05 4 629 7 676 1991 4 770 0 14957 0.0373207 0.0038322 0.10 7 639 3 861 0.26399166 0.032611216 0.10 0.05 4 590 7 603 1992 6 327 0 15443 0.0428676 0.0041637 0.10 9 872 0 4990 0.33415721 0.041899299 0.10 0.05 4 520 7 5 1 9 1993 5 288 0 15816 0.0359737 0.0034237 0.10 3 787 0.27691913 0.033524838 0.10 0.05 4 427 7 432 1994 4 309 238 13925 0.0319610 0.0032410 0.10 428 4591 0.2859991 0.033079017 0.10 0.05 4 322 7 349 1995 4 203 1086 12581 0.0434871 0.0061677 5 305 581 3 828 0.26957643 0.027277526 0.10 0.05 4 2 1 5 7 269 1996 2 909 566 6628 0.0370360 0.0044215 0.05 4 905 729 2 558 0.29824056 0.03384428 0.10 0.05 4 114 7 191 581 0.05 3 907 1997 3 319 5740 0.0462300 0.0046473 0.10 735 1597 0.27711413 0.032520406 0.10 0.05 4 026 7 110 1998 4 525 596 3844 0.0475281 0.0051195 0.10 5 202 810 948 0.2848132 0.044812013 0.10 0.05 3 953 7 023 1999 2 556 834 1591 0.0536925 0.0052808 2 881 810 707 0.23565566 0.031692771 0.10 0.05 3 894 6 9 2 6 2000 3 590 1409 3384 0.0506108 0.0053718 0.05 4 397 1 390 904 0.37973899 0.048959204 0.10 0.05 3 848 6 822 0.0058331 2001 3 673 2151 1930 0.0552146 0.05 3 492 1 649 699 0.27754683 0.034979683 0.10 0.05 3 810 6716 2002 2 747 1728 1944 0.0469249 0.0052775 0.05 3 732 1 980 816 0.34215991 0.046018169 0.10 0.05 3 776 6 6 1 6 2003 1 626 1566 1910 0.0467350 0.0049494 0.05 2 215 1 698 846 0.27413395 0.034797476 0.10 0.05 3 744 6 532 2004 3 574 2485 2262 0.0541154 0.05 5 299 3 253 725 0.34815862 0.049732368 0.05 3 709 6 472 2005 3 791 4392 3637 0.0639268 0.0065468 0.10 0.05 3 839 3 101 1074 0.25714529 0.030746628 0.10 0.05 3 673 6 442 0.0631913 2006 3 280 2604 2487 0.0071731 0.10 0.05 3 627 2 867 776 0.22514396 0.032208928 0.10 0.05 3 636 6 442 2007 3 648 5899 2530 0.0807523 0.0075121 0.10 0.05 3 954 3 989 514 0.2425778 0.027900987 0.10 0.05 3 600 6 468 2008 3 403 3325 1337 0.0746083 0.0074960 0.10 0.05 4 266 4 345 587 0.19628971 0.02567334 0.10 0.05 3 568 6511 2009 2 040 3024 1210 0.0753762 0.0068799 0.10 0.05 3 412 3 321 683 0.18147746 0.020984308 0.10 0.05 3 544 6 557 2010 3 177 4422 1930 0.0712016 0.0067978 0.10 0.05 3 313 4 458 738 0.1708923 0.020416242 0.10 0.05 3 5 3 1 6 593 2011 2 418 4088 0.0882102 0.0079853 0.10 0.05 3 505 5 196 741 0 14053484 0 013704708 0.10 0.05 3 5 2 9 6 607 788 2012 2 5 1 1 4821 728 0.0783248 0.0079611 0.10 0.05 2 775 4 577 617 0.15114891 0.016831688 0.10 0.05 3 541 6 591 2013 1531 3197 811 0.0783117 0.0073879 0.10 0.05 1 447 3 428 839 0.13909845 0.014205163 0.10 0.05 3 564 6 545 2014 933 2683 720 0.0909073 0.0073603 0.10 0.05 855 2 587 664 0.14762718 0.014975047 0.10 0.05 3 597 6 473 2015 826 2907 603 0.0648778 0.0059214 0.10 0.05 896 2 893 439 0.11178511 0.012618386 0.10 0.05 3 636 6 384 165 0.05 2016 3766 2 0.0699304 0.0071636 0.10 0.05 100 3 615 33 0.1310814 0.016983671 0.10 3 679 6 289

Faroes

Annual input data for NEAC PFA run-reconstruction & NCL models for Faroes. (Uncertainty values define uniform distribution around estimate used in Monte Carlo simulation)

Year	Catch 1SW salmon	Catch MSW salmon	Estimated % unreported catch of 1SW salmon	% uni	tainty in reported n of 1SW salmon	% wild	Natural mortality after 1st sea winter (M)
1971	2 620	105 796	1	0	5	1.0	0.03
1972	2 754	111 187	1	0	5	1.0	0.03
1973	3 121	126 012	1	0	5	1.0	0.03
1974	2 186	88 276	1	0	5	1.0	0.03
1975	2 798	112 984	1	0	5	1.0	0.03
1976	1 830	73 900	1	0	5	1.0	0.03
1977	1 291	52 112	1	0	5	1.0	0.03
1978	974	39 309	1	0	5	1.0	0.03
1979	1 736	70 082	1	0	5	1.0	0.03
1980	4 523	182 616	1	0	5	1.0	0.03
1981	7 443	300 542	1	0	5	1.0	0.03
1982	6 859	276 957	1	0	5	1.0	0.03
1983	15 861	215 349	1	0	5	1.0	0.03
1984	5 534	138 227	1	9	5	1.0	0.03
1985	378	158 103	1	0	5	0.9	0.03
1986	1 979	180 934	1	0	5	1.0	0.03
1987	90	166 244	1	0	5	1.0	0.03
1988	8 637	87 629			5	0.9	0.03
1989	1 788	121 965	1	Ó	5	0.8	0.03
1990	1 989	140 054	1	0	5	0.5	0.03
1991	943	84 935	1	0	5	0.5	0.03
1992	68	35 700	1	0	5	0.6	0.03
1993	6	30 023	1	0	5	0.7	0.03
1994	15	31 672	1	0	5	0.7	0.03
1995	18	34 662	1	0	5	0.8	0.03
1996	101	28 381	1	0	5	0.8	0.03
1997	0	0		0	0	0.0	0.03
1998	339	1 424	1	5	5	0.8	0.03
1999	0	0		0	0	0.0	0.03
2000	225	1 765	1	5	5	0.8	0.03
2001	0	0		0	0	0.0	0.03
2002	0			0	0	0.0	0.03
2003	0	0		0	0	0.0	0.03
2004	0	0		0	0	0.0	0.03
2005	0	0		0	0	0.0	0.03
2006	0	0		0	0	0.0	0.03
2007	0			0	0	0.0	0.03
2008	0			0	0	0.0	0.03
2009	0	0		0	0	0.0	0.03
2010	0	0		0	0	0.0	0.03
2011	0	0		)	0	0.0	0.03
2012	0			)	0	0.0	0.03
2013	0			)	0	0.0	0.03
2014	0	0		)	0	0.0	0.03
2015	0	0		)	0	0.0	0.03
2016	0			)	0	0.0	0.03
2017	0	0		)	0	0.0	0.03

**West Greenland** 

Annual input data for NEAC PFA run-reconstruction & NCL models for WEST GREENLAND. (Uncertainty values define uniform distribution around estimates used in Monte Carlo simulation]

Year	Declared	Estimated	Wean weight	Estimated	Estimated	Proportion	Proportion	No. Fish	No. Fish
	catch (t)	unreported		min'	max'	1SW in NAC	1SW in NEAC	identified as	identified as
		catch			proportion of	fish	fish	NAC (from	NEAC (from
				NAC fish	NAC fish			genetic	genetic
				(from scale	(from scale			analysis)	analysis)
				analysis)	analysis)				
1971	2 689.0	0.0	3.140	0.280	0.400	0.945	0.964		
1972	2 113.0	0.0	3.440	0.340	0.370	0.945	0.964		
1973	2 341.0	0.0	4.180	0.390	0.590	0.945	0.964		
1974	1 917.0	0.0	3.580	0.390	0.460	0.945	0.964		
1975	2 030.0	0.0	3.120	0.400	0.480	0.945	0.964		
1976	1 175.0	0.0	3.040	0.380	0.480	0.945	0.964		
1977	1 420.0	0.0	3.210	0.380	0.570	0.945	0.964		
1978	984.0	0.0	3.350	0.470	0.570	0.945	0.964		
1979	1 395.0	0.0	3.340	0.480	0.520	0.945	0.964		
1980	1 194.0	0.0	3.220	0.450	0.510	0.945	0.964		
1981	1 264.0	0.0	3.170	0.580	0.610	0.945	0.964		
1982	1 077.0	0.0	3.110	0.600	0.640	0.945	0.964		
1983	310.0	0.0	3.100	0.380	0.410	0.945	0.964		
1984	297.0	0.0	3.110	0.470	0.530	0.945	0.964		
1985	864.0	0.0	2.870	0.460	0.530	0.925	0.950		
1986	960.0	0.0	3.030	0.480	0.660	0.951	0.975		
1987	966.0	0.0	3.160	0.540	0.630	0.963	0.980		
1988	893.0	0.0	3.180	0.380	0.490	0.967	0.981		
1989	337.0	0.0	2.870	0.520	0.600	0.923	0.955		
1990	274.0	0.0	2.690	0.700	0.790	0.957	0.963		
1991	472.0	0.0	2.650	0.610	0.690	0.956	0.934		
1992	237.0	0.0	2.810	0.500	0.570	0.919	0.975		
1993	0.0	12.0	2.730	0.500	0.760	0.950	0.960		
1994	0.0	12.0	2.730	0.500	0.760	0.950	0.960		
1995	83.0	20.0	2.560	0.650	0.720	0.968	0.973		
1996	92.0	20.0	2.880	0.710	0.760	0.941	0.961		
1997	58.0	5.0	2.710	0.750	0.840	0.982	0.993		
1998	11.0	11.0	2.780	0.730	0.840	0.968	0.994		
1999	19.0	12.5	3.080	0.840	0.970	0.968	1.000		
2000	21.0	10.0	2.570	0.000	0.000	0.974	1.000	344	146
2001	43.0	10.0	3.000	0.670	0.710	0.982	0.978	1	1
2002	9.8	10.0	2.900	0.000	0.000	0.973	1.000	338	163
2003	12.3	10.0	3.040	0.000	0.000	0.967	0.989	1 212	567
2004	17.2	10.0	3.180	0.000	0.000	0.970	0.970	1 192	447
2005	17.3	10.0	3.310	0.000	0.000	0.924	0.967	585	182
2006	23.0	10.0	3.240	0.000	0.000	0.930	0.988	857	326
2007	24.8	10.0	2.980	0.000	0.000	0.965	0.956	917	206
2008	28.6	10.0	3.080	0.000	0.000	0.974	0.988	1 593	260
2009	28.0	10.0	3.500	0.000	0.000	0.934	0.894	1 483	138
2010	43.1	10.0	3.420	0.000	0.000	0.982	0.975	991	249
2011	27.4	10.0		0.000	0.000	0.939	0.831	888	72
2012	34.6	10.0	3.440	0.000	1.000	0.932	0.980	1 121	252
2013	47.7	10.0	3.350	0.000	1.000	0.949	0.966	938	211
2014	70.4	10.0	3.320	0.000	1.000	0.913	0.961	660	260
2015	60.9	10.0	3.370	0.000	1.000	0.970	0.982	1 337	337
2016	30.2	10.0	3.180	0.000	1.000	0.935	0.955	864	438
2017	28.0	10.0	3.490	0.000	1.000	0.925	0.931	734	252

Stock composition	
Country	MSW
France	0.027
Finland	0.001
Iceland	0.001
Ireland	0.147
Norway	0.028
Russia	0.000
Sweden	0.003
UK(England & Wales)	0.149
UK(Northern Ireland)	0.000
UK(Scotland)	0.644
Other	
Total	1.000

# Appendix 4: Input data for Atlantic salmon used to do the run-reconstruction and estimates of returns and spawners by size group and age group for North America

Appendix 4.i. Input data for the fishery at West Greenland used in the run reconstruction model

Year 1	WGHarv	WGUnHarv		WGSampleNAC	WGSampleNEAC	WGPropNACMin	WGPropNACMax		WGProp1SWNEAC
1970	0	C	3	(		0.2	0.5	0.9	1
1971	2689	C	3.14	(		0.28	0.4	0.945125	0.964125
1972	2113	C	3.44	(		0.34	0.37	0.945125	0.964125
1973	2341	C	4.18	(		0 0.39	0.59	0.945125	0.964125
1974	1917	C	3.58	(		0.39	0.46	0.945125	0.964125
1975	2030	C	3.12	(		0 0.4	0.48	0.945125	0.964125
1976	1175	C	3.04	(		0.38	0.48	0.945125	0.964125
1977	1420	C	3.2125	(		0.38	0.57	0.945125	0.96412
1978	984	C	3.35	(		0 0.47	0.57	0.945125	0.964125
1979	1395	C	3.34	(		0.48	0.52	0.945125	0.96412
1980	1194	C	3.22	(		0.45	0.51	0.945125	0.964125
1981	1264	C	3.17	(		0.58	0.61	0.945125	0.964125
1982	1077	C	3.11	(		0.6	0.64	0.945125	0.964125
1983	310	C	3.1	(		0.38	0.41	0.945125	0.964125
1984	297	C	3.11	(		0 0.47	0.53	0.945125	0.964125
1985	864	C	2.87	(		0.46	0.53	0.925	0.95
1986	960	C	3.03	(		0.48	0.66	0.951	0.975
1987	966	C	3.16	(		0 0.54	0.63	0.963	0.98
1988	893	C	3.18	(		0 0.38	0.49	0.967	0.983
1989	337	C	2.87	(		0 0.52	0.6	0.923	0.955
1990	274	C	2.69	(		0 0.7	0.79	0.957	0.963
1991	472	C	2.65	(		0 0.61	0.69	0.956	0.934
1992	237	C	2.81	(		0 0.5	0.57	0.919	0.975
1993	0	12	2.725	(		0 0.5	0.76	0.946	0.96075
1994	0	12	2.725	(		0.5	0.76	0.946	0.96075
1995	83	20	2.56	(		0 0.65	0.72	0.968	0.97
1996	92	20	2.88	(		0 0.71	0.76	0.941	0.963
1997	58	5	2.71	(		0 0.75	0.84	0.982	0.993
1998	11	11	2.78	(		0 0.73	0.84	0.968	0.994
1999	19	12.5	3.08	(		0.84	0.97	0.968	
2000	21	10	2.57	344	14	6 0	C	0.974	
2001	43	10	3			1 0.67	0.71	0.982	0.978
2002	9.8	10	2.9	338	16	3 0		0.973	
2003	12.3	10	3.04	1217	56	7		0.967	0.989
2004	17.2	10	3.18	1192	44	7 (		0.97	0.97
2005	17.3	10	3.31	585	18	2 0	C	0.924	0.96
2006	23	10	3.24	857	32	6 0	C	0.93	0.988
2007	24.8	10	2.98	917	20	6 0	C	0.965	0.956
2008	28.6	10	3.08	159	26	0 0	C	0.974	0.988
2009	28	10	3.5			8 0	C	0.934	0.894
2010	43.1			991			C		
2011	27.4						C		
2012	34.5			1121			C		
2013	47.7			938					0.966
2014	70.482	_ 10		660					
2015	60.9			1337					0.982
2016	30.249			864					0.955
2017	28		_	734					0.931

Appendix 4.ii. Input data for sea fisheries on large salmon and small salmon from Newfoundland and Labrador used in the run reconstruction model. Labrador represents harvests from Labrador in aboriginal fisheries for food, social and ceremonial purposes and the resident food fishery beginning in 1998.

0 81152 43041 85904 73961 100504 79318	0 0 42861				LB_SFA2_Lg_Comm					LB_SFA14B_Sm_Comm	
43041 85904 73961 100504 79318					45479	9595	0		29441	8605	
85904 73961 100504 79318	42861				64806		0		38359	11212	
73961 100504 79318	40.007		111141				0		28711	8392	
100504 79318	43627 85714		176907 153278	30204 13866			0		6282 37145	1836 9328	
79318	72814		91935				0		57560		
	95714		118779				0		47468		
114413	63449		57472				0		40539	11267	
64073	37653		38180				0		12535	4026	
29936	29122	140844	62622	21291	27073	3874	0	9712	28808	7194	
86941	54307	186648	94291	28750	87067	9138	0	22501	72485	8493	
98672	38663		60668				0		86426		
37576									34200		
31847									20699	3592	
25792									20055	5303	
0									13336		
0	0	0	0	3825	9651	3620	0	3893	12037	1144	
0									4535	802	
							0				
							0				
						0					
							-				
0						0					
0	0	0	0	0	0	0	3696	0	0	0	838
					0						
0	0				0						
0				0	0						
0	0	0	0	0	0	0	6208	0	0	0	690
	460776 48218 44540 48218 44540 48218 44540 48218 48296 67072 36449 67072 48250 67072	46076 35055 46276 35055 48218 28215 44540 15135 36975 24383 36975 24383 36975 15577 31847 11639 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	46076   35055   14348   48218   28215   116592   44540   15135   88184   36975   24383   131360   57072   19241   192208   36449   14763   15375   31847   11639   71761   7	46076   35055   144445   77017     48218   28215   116592   55683     48450   15135   98184   52813     36975   24883   131360   79275     67072   19241   192308   82401     67072   19241   192308   82401     67072   19241   192308   82401     67072   19241   192308   82401     67072   19241   192308   82401     67072   19241   192308   82201     67072   19241   192308   82401     74761   46053     74761   46053     7577   74620   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     76084   74761     7		46076   35055   143445   77017   24192   53085   48218   28215   116592   55683   19403   33320   48218   28215   116592   55683   19403   33320   36975   24383   131360   79275   13252   16789   48896   22036   151275   91912   19152   34071   67072   19241   192308   82401   18257   49799   36449   14763   115375   60884   16661   26836   31847   11639   71761   46053   7313   17316   25792   10259   62331   42721   1369   7579   25792   10259   62331   42721   1369   7579   0		66076   35055   143445   77017   24192   53085   5966   0		66076   35055   143445   77017   24192   53085   5966   0   18478   53592     48218   28215   116592   55683   19403   33320   7489   0   15964   30185     48218   28215   116592   55683   19403   33320   7489   0   15964   30185     48218   28215   116592   55683   19403   33320   7489   0   11474   11695     36975   24383   131360   79275   13252   16789   3994   0   15400   24499     48996   22036   151275   91912   19152   34071   5342   0   17779   45321     67072   19241   192308   82401   18257   49799   11114   0   13744   64351     36449   14763   115375   74620   12621   3286   4591   0   19641   56381     37576   15577   116375   60884   15651   26836   4646   0   13233   34200     31847   11639   77761   46053   7313   17316   2858   0   8736   20699     25792   10259   62331   42721   1369   7679   44417   0   1410   20055     0	46076   35055

Appendix 4.iii. Input data for sea fisheries on large salmon and small salmon from St Pierre & Miquelon used in the run-reconstruction model.

Year of the fishery	Reported harvest (kg)	Professional (kg)	Recreational (kg)	Number of salmon (prof)	Number of salmon (recr)	Mean weight (prof)	Mean weight (recr)		large salmon		salmon	Weight sampled (gutted)	Mean weight (gutted)	Number sampled for weight
1970	0							0	0	0				
1971	0							0	0					
1972	0							0	0					
1973	0							0	0	0				
1974	0							0	0	0	2.779			-
1975	0							0	0	0			_	
1976	3000							1080	348	731	2.779			
1977	0							0		0				
1978	0							-	0	0				
1979	0							0						
1980	0							0		0				
1981	0							0		0				
1982								-	0		2.779			
1983 1984	3000 3000							1080 1080	348	731	2.779 2.779			
									348	731				
1985	3000							1080	348	731	2.779			
1986 1987	2500 2000							900	290 232	609 487	2.779			
								720			2.779			
1988	2000							720	232	487	2.779			
1989 1990	2000 1880	1146	734					720 677	232 218	487 458	2.779 2.779			
1990	1162	632	530					418	135	283			_	
1991	2319	1295						834	269	565	2.779			
1992	2943	1902	1024 1041					1059	342	717	2.779			
1993	3423	2633	790					1232	398	834	2.779			+
1995	837	392	445					301	97	204	2.779			+
1996	1568	951	617					564	182	382				+
1996	1491	762	729					537	173	363	2.779			
1997	2307	1039	1268				-	830	268	562	2.779			
1998	2322	1182	1200					836	270	562			_	
2000	2267	1134	1133					816		552				
2000	2155	1544	611					775	250	525				
2001	1952	1223	729					702	230	476				
2002	2892	1620	1272		530	2.949	2,402	1079	348	731	2.680		2 2.50	6 340
2003	2784	1499	1285		535	2.711	2.402	1088	196	892	2.559			
2005	3287	2243	1044		435	2,664	2,402	1277	351	926	2.574			
2005	3555	1730	1825		819	2.724	2.228	1454	469	985	2.445			
2007	2032	970	1062		470	4.696	2,260	677	218	458	3.003			
2007	3450	1604	1846		933	3.687	1.979	1368	442	926	2.522			
2009	3464	1864	1600		748	3.603	2.139	1265	408	857	2.737		3.2	1 00
2010	2782	1002	1780		768	3.289	2.318	1073	470	602	2.737		3 2.86	6 57
2010	3756	1764	1992		819	4.947	2.432	1176	1031	145	3.195			
2012	1446	278	1168		405	3,603	2.884	482	156	327	2.999		4.3	10
2012	5300	2290	3010		1253	4.083	2.402	1814	1272	542			3.55	5 60
2013	3811	2250	1561	526	525	4.278	2.402	1051	611	440	3.626			
2014	3510	1210	2300		958	2.747	2.402	1398	410	988	2.510			
2015	4728	979	3749		1246	2.245	3.009	1682	286	1396	2.811			
		9/9	3/49	430	1240	- 2.245	3.009	1002	200	1390	2.011	290	2.04	140

<sup>\*</sup> whole weight is assumed to be gutted weight sampled X 1.15.

Appendix 4.iv. Input data for large salmon for Labrador used in the run-reconstruction.

	IR SFA1	IR SFA2	LB SFA14												Ret River	Ret.River		
		Lg_Com	-		NR SEΔ1	pLB_SFA1 p	IR SFA2	nIR SEA2	nIR SEA1	nIR SFA1	FR IR Iσ	FR IR Iø					LB_Ang_L	ΙΒ Δησ
		m	mm		Lg L	. – .	Lg L	-		4B_Lg_H			p2SW L	p2SW_H		H		g_Rel
1970	17633	45479		-	0.6	0.8	0.6				_	0.9	0.7	0.9	. 0			
1971	25127	64806			0.6	0.8	0.6					0.9	0.7	0.9	0			
1972	21599	55708			0.6	0.8	0.6					0.9	0.7	0.9	0			
1973	30204	77902			0.6	0.8	0.6	0.8				0.9	0.7	0.9	0			
1974	13866	93036		-	0.6	0.8	0.6	0.8			-	0.9	0.7	0.9	0			
1975	28601	71168			0.6	0.8	0.6	0.8					0.7	0.9	0			
1976	38555	77796			0.6	0.8	0.6	0.8				0.9		0.9	0			
1977	28158	70158			0.6	0.8	0.6					0.9	0.7	0.9	0			
1978	30824	48934			0.6	0.8	0.6						_	0.9	0			
1979	21291	27073			0.6	0.8	0.6						0.7	0.9	0			
1980	28750	87067			0.6	0.8	0.6				_		0.7	0.9	0			
1980	36147				0.6	0.8	0.6						0.7	0.9	0			
1981			_	-	0.6	0.8						0.9	0.7		0			
	24192			-			0.6		-				-	0.9	0	_	-	
1983	19403	33320			0.6	0.8	0.6					0.9 0.9	0.7 0.7	0.9	0			
1984	11726				0.6	0.8	0.6		_					0.9				
1985	13252	16789		-	0.6	0.8	0.6						0.7 0.7	0.9	0			
1986	19152				0.6	0.8	0.6	_						0.9	-		-	
1987	18257	49799			0.6	0.8	0.6	_					0.7	0.9	0			
1988	12621	32386			0.6	0.8	0.6	_					0.7	0.9	0			
1989	16261		_		0.6	0.8	0.6						0.7	0.9	0			
1990	7313			-	0.6	0.8	0.6						0.7	0.9	0	_		
1991	1369	7679		-	0.6	0.8	0.6	0.8	A 7			0.9	0.7	0.9	0			
1992	9981	19608			0.6	0.8	0.6				0.580286		0.7	0.9	0			
1993	3825	9651			0.6	0.8	0.6				0.381506		0.7	0.9	0			
1994	3464	11056			0.6	0.8	0.6	0.8			0.293168		0.7	0.9	0			
1995	2150	8714	-	-	0.6		0.6	0.8				0.251348	0.7	0.9	0			50
1996	1375	5479			0.6	0.8	0.6	0.8			0.126963		0.7	0.9	0			
1997	1393	5550			0.6433	0.7247	0.8839	0.9521			0.17	0.3	0.7	0.9	0			
1998	0			-	1	1	1					0.3	0.6	0.71	7374			
1999	0				1	1	1	1		_		0.3	0.6	0.71	8827			
2000	0				1	_	1	1				0.3	0.6	0.71	12052			
2001	0				1	1	1					0.3	0.6	0.71	12744			
2002	0				1	1	1					0.3	0.6	0.71	9076			9
2003	0			_	1	1	1	_	_	_		0.3	0.6	0.71	6676			
2004	0				1	1	1					0.3	0.6	0.71	10964			_
2005	0			-	1	1	1		_			0.3	0.6	0.71	11159			
2006	0				1	1	1					0.3	0.6	0.71	12414			
2007	0				1	1	1	_	_	_			0.6	0.71	11887			
2008	0				1	1	1			_			0.6	0.71	14700			
2009	0				1	1	1						0.6	0.7	18643			
2010	0				1	1	1						0.6	0.7	10764			
2011	0				1	1	1						0.6	0.7	30198			
2012	0			4228	1	1	1	1	_	_		0.3	0.6	0.7	19062	48538	0	10
2013	0				1	1	1	1				0.3	0.6	0.7	36859			
2014	0	C	0	3994	1	1	1	1	1	1	0.17	0.3	0.6	0.7	36055	87989	0	160
2015	0	C	0	6146	1	1	1	1	1	1	0.17	0.3	0.6	0.7	49662	127898	0	130
2016	0	C	0	5595	1	1	1	1	1	1	0.17	0.3	0.6	0.7	36134	108273	0	320
2017	0	C	0	6208	1	1	1	1	1	1	0.17	0.3	0.6	0.7	32055	121307	0	253

Appendix 4.v. Input data for small salmon for Labrador used in the run-reconstruction.

	LB SFA1	LB SFA2	LB SFA14										Ret.River	Ret.River		
			_	NSm LBF	pLB SFA1	pLB_SFA1	nLB SFA2	pLB SFA2	pLB SFA1	pLB SFA1	ER LB S	ER LB S			LB Ang S	LB Ang S
Year	m	m	mm	SC	Sm L	Sm H	Sm L	Sm H		4B Sm H		m H	L	Н		m Rel
1970	14666		8605	0			0.6					0.5		_		0
1971	19109		11212	0			0.6					0.5				0
1972	14303	28711	8392	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	2947	0
1973	3130			0	0.6		0.6				0.3	0.5		0	7492	0
1974	9848	37145	9328	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	2501	0
1975	34937	57560	19294	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	3972	0
1976	17589	47468	13152	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	5726	0
1977	17796	40539	11267	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	4594	0
1978	17095	12535	4026	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	2691	0
1979	9712	28808	7194	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	4118	0
1980	22501	72485	8493	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	3800	0
1981	21596	86426	6658	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	5191	0
1982	18478	53592	7379	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	4104	0
1983	15964	30185	3292	0	0.6	0.8	0.6			0.8	0.3	0.5	0	0	4372	0
1984	11474	11695	2421	0	0.6	0.8	0.6	0,8	0.6	0.8	0.3	0.5	0	0	2935	0
1985	15400	24499	7460	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	3101	0
1986	17779	45321	8296	0	0.6	0.8	0.6	0.8	0.6	0.8	0.3	0.5	0	0	3464	0
1987	13714					0.8	0.6	0.8	0.6							0
1988	19641						0.6									0
1989	13233						0.6									0
1990	8736						0.6									0
1991	1410							_								
1992	9588						0.6	_				0.392801				
1993	3893					$\overline{}$	0.6	_				0.245508				1793
1994	3303			0			0.6					0.185922				3681
1995	3202				_		0.6									3302
1996	1676						0.6					0.06772				3776
1997	1728				0.3557		0.748					0.082				2187
1998	0					1	1					0.082				3758
1999	0				1	-	1					0.082				4407
2000	0		_		1		1					0.082				7095
2001	0				1		1					0.082				4640
2002	0		_		1							0.082				5052
2003	0	_					1					0.082				4924
2004	0				1							0.082				5968
2005	0						1									7120
2006	0		_	-			1					0.082				
2007	0															
2008	0				1											
2009	0															3396
2010	0				1											4704
2011	0															
2012	0															
2013	0															4167
2014	0															
2015	0															
2016	0															
2017	0	0	0	6907	1	1	1	1	1	1	0.045	0.082	80484	246247	1455	3265

Appendix 4.vi. Input data for returns of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

		Salmon Fishi	ng Area 3			Salmon Fish	hing Area 4			Salmon Fis	hing Area 5			Salmon F	ishing Area	6		Salmon	Fishing A	ea 7		Salmon	Fishing Area	a 8
	Small	salmon	Large	salmon	Smal	l salmon	Large	salmon	Small	salmon	Large	salmon	Smal	l salmon	Large	salmon	n Sm	all salmon	Lai	ge salmon	Smal	l salmon	Lar	ge salmon
	Returns		Returns		Returns		Returns		Returns		Returns		Returns		Returns		Return	ıs	Return	ıs	Returns		Returns	s
Year	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
ugs labe	SFA3Sm_L[]	SFA3Sm_H[]	SFA3Lg_L[]	SFA3Lg_H	[] SFA4Sm_L[]	SFA4Sm_H[]	SFA4Lg_L[]	SFA4Lg_H[]	SFA5Sm_L[]	SFA5Sm_H[]	SFA5Lg_L[]	SFA5Lg_H[]	SFA6Sm_	L[SFA6Sm	_H SFA6Lg_I	L[SFA6L	_g_H SFA7S	m_l SFA7S	m_l SFA7L	g_L SFA7Lg	_H SFA8Sm_	L[ SFA8Sn	n_H∣SFA8Lç	_L  SFA8Lg_
1970			155	7.	37 1610							209			560 1	7	79	67	133	4		52	123	4 1
1971			146		98 126											1	52	133	267			33	167	5 2
1972					68 114											13	112	203	407			93	187	6 2
1973																19	235	437	873				627	19 8
1974					45 179										)20 11		233	443	887				340	20 3
1975					41 198											14	88	133	267				580	41 8
1976					60 222								_	3 10		19	198	100	200		24 20		533	32 6
1977					64 279									7 20	573 12		253	260	520				540	26 5
1978					99 292											8	117	330	660				293	9 1
1979					79 267:											17	93	417	833				667	19 3
1980					22 3138											18	96	340	680				800	18 3
1981													201		)33 26		536	410	820				513	34 6
1982					24 3324											:3	46		1033				567	7 1
1983					95 2984								1 98			15	91	463	927				273	6 1
1984					40 3493							145	7 110			7	168	339	722				594	4 4
1985					83 4440							181	6 156			4	224	408	845				777	6 5
1986					13 340						_	5 217				10	238	373	779				1054	12 7
1987					18 214								9 54			4	77	110	222				340	4 2
1988					56 371						_	7 226				19	234	483	995				614	7 4
1989					61 1540							88	5 100			30	154	269	547	-		)3	820	12 6
1990					20 2224							129	3 131			6	179	193	337				591	12 4
1991			193		50 2100					_						.7	106	155	254			17	78	2
1992											516 586					51	603	292	585			0	0	0
1993 1994					14 456											1 2	352 135	462 64	890				813	15 5 7 2
1994			609		68 2940 65 3143											3	101	233	141 560				243 446	7 2 11 4
1995									_		_					17	141	151					500	16 4
1990			1226							38875 11786						15	79	60	338 110				110	6 2
1997			1956													i4	192	249	350				227	16 5
1999									1007	160/2	964					17	122	69	100				220	14 4
2000								_	12414	16734						36	218	159	214			)6	143	8 2
2000																3	53	53	69	4		20	26	2 2
2001																.6	46	0	09	0		72	108	5 1
2002																13	74	104	126			52	63	3
2003					86 4959			1046	8385							9	38	0	0	0		11	64	2
2004								15329							201	9	48	0	0	0	0	26	76	2 1
2006															200	8	21	0	0	0			275	20 5
2007					28 3693				8734							9	28	129	262			17	35	20 3
2008								15508	11459							15	81	84	141				329	15 4
2000							3518	22760								.9	185	0	0	0			252	8 5
2010							6094				1766				43 7				263	0	46 11		137	8 2
2011			1308												48 7				170		40 27			23 10
2012															74 5				159		21 40			29 7
2012															67 4				429		27 12		185	7 5
2013			1132												23 4		110		208		28 10		163	8 2
2015			1709													2	236		228		34 33			28 8
2016			1129												09 4		127		236		33 22			19 6
2017																1	29	36	64	3	8 13			11 2

Appendix 4.vi. (Continued). Input data for returns of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

abic. O	mall and larg	je returns io	i ive wiouilui	and.																				
			shing Area 9				ning Area 10			Salmon Fish					shing Area 12			Salmon Fish				Salmon Fish		
		salmon		salmon		salmon	Large s	almon		salmon		salmon		salmon		salmon		salmon		salmon		salmon		salmon
V	Returns Min	Max	Returns Min	Mari	Returns Min	Max	Returns Min	Ma	Returns Min	Max	Returns Min	Mari	Returns Min	Max	Returns	Mari	Returns Min	Max	Returns	Max	Returns Min	Max	Returns Min	Man
				Max				Max				Max				Max	H SFA13Sm_L[]							Max
1970								565 565																
1970								872								00 42					12523			
1972								533													8057			
1973								1678													17607			
1974								932								53 50					10400			
1975								401													16060			
1976								586							80 11						24603		1437	
1977								691													19023			
1978								629								2 14					10803			
1979								324								0 28					21927			
1980								456													12477			
1981								1167							67 12 67 47	0 93								
1982								202													15877			
1983	7677	1535	3 35	2 705	5 380	0 760	0 174	349		31793	730	1460	2227	3 44	47 10	)2 20	14 20828	31701	4465	5 5100	12667	25333	1704	
1984	7989	1702	3 12	5 1221	1 514	1 1095	5 81	785	24767	52774	389	3784			51 10	06 103	36 26184	4 37852	2296	6 3710	16962	36143	266	259
1985	6375	1319	8 9	6 912	2 483	1 1000	0 73	691	21213	43914	320	3034	3996	82	73 6	50 57	2 16028	3 25505	1375	5 2508	13209	27345	199	
1986	8411	1755	5 20	8 123	1 5619	9 1172	7 139	822	20300	42368	501	2970	3433	71	66 8	35 50	)2 22881	36916	2079	3649	18411	38426	455	269
1987	3416	686	5 8	8 489	9 169	0 339	7 44	242	15087	30317	391	2162	3274	1 65	80 8	35 46	59 19629	32325	1546	6 3022	18203	36580	471	260
1988	5179	1066	8 12	4 748	8 430	8 887	3 103	622	18985	39106	456	2741	5330	109	79 12	28 77	70 26162	2 43480	1950	3917	23580	48570	566	340
1989	5352	1089	5 16	0 82	1 365:	5 744	0 109	561	12047	24524	360	1849	2279	46	40 6	58 35	50 10154	16156	849	9 1565	13036	26537	390	200
1990	7332	1283	4 25	6 1000	328	1 574	3 115	447	17470	30578	610				87 11	17 45	59 21518	31183	1778	8 3084	19843	34732	693	
1991								131								95 36				9 2433	15307	25141		
1992								642			605		4671								34927		1271	
1993	11402	2194						763			865							3 46889	2618		31116	59893	1095	
1994								635					2761								13321		783	
1995								1142					2294								20840			
1996								1805																
1997								1712			1735										25241			
1998								2788				3018									23995		2392	
1999								1535			887										26960			
2000								1475			1310													
2001								523													20775			
2002								709													26558		1729	
2003								608			718										40802			
2004								780		21443											30057			
2005								1159													17340			
2006								1609	11033												28081		3266 2264	
2007 2008								1285	2020	11449 18654											19966 25802			
2008								1109	11136 7536												23802			
2009								1783													31675			
2010								3600													24110			
2011								1579													35229		2485	
2012								3527	7372												13924			
2013								826													23640		1851	
2015								1184													32384		2727	
2016								1312													22013		1876	
2017								611													21179		1704	

Appendix 4.vii. Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

		Salmon Fishir	g Area 3			Salmon Fis	hing Area	4		Salmon Fi	shing Area	1 5		Salmon F	ishing Area	a 6		Salmon Fi	shing Area	a 7		Salmon F	ishing Area	a 8
	Smal	Isalmon	1	salmon		salmon		salmon		salmon		salmon	Smal	l salmon		salmon	Smal	l salmon	_	salmon		salmon		salmon
	Spawners		Spawners	s	Spawners		Spawner	s	Spawners	s	Spawner	rs	Spawne	rs	Spawne	rs	Spawne	rs	Spawner	rs	Spawner	s	Spawner	rs
Year	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
	SFA3SSm_L	] SFA3SSm_H[	] SFA3SLg	SFA3SLg	_SFA4SSm_	SFA4SSm	SFA4SLg	SFA4SLg	SFA5SSm	SFA5SSm	SFA5SLg	_SFA5SLg	SFA6SS	m SFA6SSI	m SFA6SL	SFA6SLg	SFA7SS	m SFA7SSr	n SFA7SLg	SFA7SLg	SFA8SSr	n SFA8SSi	m SFA8SLg	_SFA8SLc
1970	1829	4443	154	736	11314	27478	910	4512	5194	12614	404	2058	196	476	14	76	47	113	3	18	43	105	0	13
1971	1731	4205	135	687	8827	21437	688	3499	3920	9520	293	1541	128	312	10	51	93	227	8	38	58	142	0	15
1972	1162	2822	98	468	8036	19516	655	3214	4422	10738	354	1762	278	674	23	112	142	346	12	57	65	159	6	26
1973	2772	6732	232	1115	15657	38023	1275	6259	4928	11968	405	1974	583	1417	49	235	306	742	26	123	219	533	15	84
1974	1958	4754	318	641	12537	30447	1983	4049	3820	9276	608	1237	707	1717	115	232	310	754	49	100	119	289	20	39
1975	2583	6273	520	1041	13867	33677	2628	5421	4639	11265	912	1846	219	533	43	87	93	227	19	38	203	493	41	82
1976	2210	5366	379	759	15594	37870	2495	5177	4429	10755	697	1459	576	1400	97	196	70	170	12	24	187	453	32	64
1977	3570	8670	478	960	19591	47577	1559	4204	10771	26157	1410	2864	936	2272	107	234	182	442	24	48	189	459	26	51
1978	1769	4295	149	298	20473	49719	1229	2959	6669	16195	536	1100	691	1677	51	110	231	561	19	38	103	249	9	17
1979	4760	11560	390	779	18727	45481	1206	2740	3106	7542	234	489	569	1383	45	91	292	708	24	48	233	567	19	38
1980	4067	9877	224	485	21966	53346	903	2312	6305	15311	376	780	747	1813	34	82	238	578	14	30	280	680	18	36
1981	5502	13362	1042	2087	31584	76704	5637	11635	8139	19765	1511	3056	1412	3428	239	507	287	697	53	107	180	436	34	68
1982	6146	14926	124	336	23270	56514	544	1346	5677	13787	143	338	672	1632	6	29	362	878	2	15	198	482	0	5
1983	3773	9163	245	493	20893	50739	1073	2443	5500	13356	191	551	691	1677	35	81	324	788	0	9	96	232	1	8
1984	2531	6525	55	540	25033	64536	533	5322	6835	17620	149	1456	789	2034	12	163	243	626	1	48	200	515	4	43
1985	3462	8569	72	683	32218	79741	671	6352	9208	22791	192	1816	1134	2806	24	224	296	733	6	58	272	674	6	54
1986	2054	5126	70	413	24722	61700	840	4977	10782	26910	366	2170	1184	2955	40	238	271	677	9	55	367	916	12	74
1987	1655	3895	57	318	16032	37722	556	3078	4892	11511	170	939	403	948	14	77	82	194	3	16	126	297	4	24
1988	4868	11888	159	956	27317	66712	892	5367	11549	28204	377	2269	1189	2904	39	234	355	867	12	70	219	535	7	43
1989	2266	5376	90	461	11623	27581	461	2365	4350	10323	172	885	755	1792	30	154	203	481	8	41	304	721	12	62
1990	5032	10098	236	920	16583	33273	776	3033	7071	14188	331	1293	978	1963	46	179	144	288	7	26	252	505	12	46
1991	4334	7965	193	750	16113	29607	718	2788	6745	12395	301	1167	613	1126	27	106	119	218	5	21	36	67	2	6
1992	9844	21262	415	4094	33228	71898	1407	13866	12175	26363	516	5088	1450	3132	61	603	252	545	11	105	0	0	0	0
1993	10054	20961	400	1599	39162	81344	1590	6226	14370	29779	576	2270	2243	4623	90	351	404	831	16	63	369	760	15	58
1994	9146	24802	749	3247	20576	55760	1644	7259	6855	18510	560	2420	381	1026	30	133	46	122	4	16	79	212	6	27
1995	7409	21791	580	2636	22872	67179	1801	8135	8122	23776	642	2880	287	831	23	100	173	501	14	60	135	397	11	48
1996	15729	39860	1412	4247	42346	107268	3757	11383	14095	35586	1263	3787	522	1317	46	139	124	311	11	33	180	457	16	48
1997	9422	19095	1209	3954	19309	39107	2467	8083	5228	10547	668	2177	190	384	24	79	49	99	6	20	48	98	6	20
1998	16390	24345	1933	6969	43559	64785	5160	18599	9943	14753	1155	4201	455	673	53	191	212	313	25	88	135	201	16	57
1999	11804	18173	1279	4189	52390	80698	5650	18583	8832	13603	947	3126	343	528	37	121	58	90	6	21	119	188	14	45
2000	17003	23723	1449	3711	32879	45946	2803	7201	10897	15217	923	2377	993	1386	84	217	140	195	12	31	88	125	8	20
2001	9861	13489	892	2089	33365	45682	3023	7086	8344	11433	767	1786	250	342	23	53	42	59	4	9	17	23	2	4
2002	8620	13856	671	1994	28099	45208	2175	6498	3194	5124	250	737	199	319	15	45	0	0	0	0	55	91	5	14
2003	19386	23938	1085	3478	67026	82739	3738	12001	5926	7312	331	1060	412	508	23	74	94	116	5	17	47	58	3	8
2004	6942	11402	390	1680	43104	70785	2430	10438	7307	11987	412	1766	158	259	9	38	0	0	0	0	35	58	2	9
2005	5056	17534	473	2664	28896	100323	2695	15235	4200	14518	394	2205	92	314	8	47	0	0	0	0	18	69	2	11
2006	9402	15839	1228	3216	37156	62732	4925	12825	7495	12623	969	2554	61	102	8	20	0	0	0	0	141	244	20	52
2007	9147	19842	1171	3818	32243	70143	4122	13501	7641	16603	978	3196	68	148	8	28	112	245	12	45	15	33	2	6
2008	11799	21183	1045	3379	53591	96443	4745	15402	9669	17405	867	2791	274	497	22	78	69	125	4	18	159	292	15	48
2009	11205	22795	779	5080	49881	101728	3491	22732	8828	18065	622	4049	412	834	28	185	0	0	0	0	111	228	7	51
2010	18364	23569	1595	4581	69075	88772	6006	17304	20114	25822	1754	5028	874	1120	76	217	183	235	16	46	93	120	8	24
2011	13193	24264	1291	6261	50806	93428	4789	23920	12075	22230	1176	5734	716	1314	70	339	83	153	8	39	220	412	22	108
2012	21149	31048	1639	4394	64959	95436	5046	13528	14554	21377	1140	3039	738	1086	57	154	100	147	8	21	361	533	25	73
2013	7822	12219	495	4039	44838	70136	2889	23284	15027	23531	976	7833	721	1128	44	373	241	379	16	127	102	162	7	55
2014	11917	19224	1112	2885	35599	57397	3306	8596	7318	11822	680	1773	452	728	39	106	106	176	8	25	87	142	8	22
2015	17382	29684	1679	4862	69539	118627	6699	19404	14503	24765	1399	4055	842	1435	81	234	117	203	12	34	280	484	28	81
2016	10280	22007	1078	3438	32370	69605	3473	10969	6885	14754	733	2317	373	799	41	127	92	202	10	32	174	375	19	59
201	7 13137	25841	1236	3385	27963	55033	2681	7260	4048	7985	390	1056	103	212	11	29	27	55	1	5	108	217	11	29

Appendix 4.vii. (Continued). Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

	8	Salmon Fis	shing Area	9	8	Salmon Fis	shing Area	10	9	almon Fis	hing Area	11	8	Salmon Fis	shing Area	12	5	Salmon Fish	ning Area 1	3		Salmon Fish	hing Area 14	4
	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Small	salmon	Large	salmon	Smal	salmon	Large	salmon
	Spawners	i	Spawner	S	Spawners	S	Spawner	s	Spawners	i	Spawne	rs	Spawners	s	Spawner	s	Spawners		Spawner	s	Spawners		Spawners	
Year	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
⊣[]	SFA9SSm	SFA9SSm	SFA9SLg	SFA9SLg	SFA10SS		Sn SFA10SL	g SFA10SL		SFA11SS	n SFA11SL		g SFA12SS	n SFA12SS	n SFA12SL	g SFA12SL	.g SFA13SSm		n SFA13SL		SFA14ASSn	SFA14ASSm	_ SFA14ASLg	_SFA14ASLg
1970	4417	10727	361	1768	1402	3406	112	558	11732	28492	918	4653	1748	4244	69	625	16203	28543	1608	3417	10372	25188	134	2340
1971	3780	9180	301	1504	2165	5259	166	855	9473	23007	736	3752	1059	2573	74	411	16489	30629	1633	3705	8766	21290	0	1850
1972	2658	6454	217	1063	1323	3213	108	529	11445	27795	882	4525	2165	5259	163	852	15125	29188	2004	4066	5640	13696	83	1283
1973	5040	12240	406	2011	4165	10115	310	1636	11331	27517	923	4530	1507	3661	102	582	17019	29959	1911	3808	12325	29931	91	2713
1974	3486	8466	565	1140	2828	6868	452	918	10444	25364	1682	3403	1535	3729	240	493	12085	21635	1997	3341	7280	17680	789	1692
1975	4368	10608	874	1754	996	2420	192	392	10502	25506	2076	4192	1190	2890	220	459	21668	42421	3611	6538	11242	27302	417	925
1976	3787	9197	639	1291	1703	4137	283	576	9716	23596	1629	3301	693	1683	114	233	18999	36519	2752	4862	17222	41826	1337	2774
1977	2520	6120	331	671	2560	6216	341	686	9557	23211	1272	2563	1302	3162	128	304	10898	18528	1828	2549	13316	32340	194	859
1978	3040	7384	240	497	3722	9038	273	587	9324	22644	770	1558	854	2074	52	124	12518	22392	3861	4434	7562	18366	194	460
1979	3976	9656	311	636	1981	4811	154	316	8003	19437	648	1304	1710	4154	130	270	14363	25820	1070	1749	15349	37275	174	408
1980	5551	13481	295	651	3556	8636	201	429	11828	28724	715	1474	1913	4647	94	217	18625	30958	4243	4920	8734	21210	514	1208
1981	4345	10551	773	1598	3073	7463	555	1138	16478	40018	3088	6217	2473	6007	453	922	22059	36689	4485	6789	13725	33331	953	2043
1982	4258	10342	114	260	2931	7117	91	192	17122	41582	537	1127	3628	8812	110	235	22062	37132	2847	3236	11114	26990	2987	6082
1983	5374	13050	281	634	2660	6460	95	270	11128	27024	703	1433	1556	3780	94	196	14491	25364	3855	4490	8867	21533	1635	3338
1984	5725	14759	120	1216	3684	9498	79	783	17748	45755	374	3769	4860	12529	38	968	18413	30081	1987	3401	12155	31336	179	2504
1985	4625	11448	96	912	3505	8674	73	691	15390	38091	320	3034	2899	7176	57	569	10726	20203	1349	2482	9583	23719	197	1887
1986	6113	15257	208	1231	4084	10192	139	822	14754	36822	501	2970	2495	6228	81	499	15535	29570	2013	3583	13381	33396	445	2683
1987	2549	5998	88	489	1261	2968	44	242	11258	26488	391	2162	2443	5749	82	466	13611	26307	1512	2988	13583	31960	467	2604
1988	3806	9295	124	748	3166	7731	103	622	13952	34073	456	2741	3917	9566	126	767	17945	35263	1909	3877	17329	42319	549	3388
1989	4037	9580	160	821	2757	6542	109	561	9087	21564	360	1849	1719	4080	67	349	6980	12982	836	1552	9833	23334	385	1996
1990	5466	10968	256	1000	2446	4908	115	447	13024	26132	610	2382	2507	5031	114	456	14866	24531	1744	3051	14793	29682	679	2692
1991	1844	3389	82	319	758	1392	34	131	6103	11215	272	1056	2121	3898	93	365	11037	15757	1689	2413	11742	21576	512	2020
1992	4334	9378	183	1809	1496	3287	65	642	14239	30854	605	5958	3985	8657	162	1667	20506	38635	2992	8833	30096	65023	1234	12488
1993	9956	20502	400	1559	4809	9967	194	761	21423	44150	861	3359	5176	10666	207	810	22341	41708	2544	4673	27010	55787	1058	4221
1994	2124	5723	172	746	1804	4848	144	630	5295	14449	430	1891	1949	5253	154	681	15381	30444	3207	5611	9385	25327	742	3286
1995	3887	11386	304	1376	3218	9378	253	1133	7770	22930	625	2792	1689	4922	130	592	20570	43329	1607	4860	15218	44587	1187	5385
1996	4868	12304	431	1304	6687	16880	592	1789	15226	38638	1362	4113	4082	10295	358	1088	29056	60152	3199	6852	26584	67085	2357	7115
1997	2927	5918	372	1221	4086	8257	519	1702	13304	26995	1718	5602	3655	7401	464	1527	25508	40599	3985	8266	20359	41117	2578	8467
1998	3937	5840	458	1663	6606	9777	771	2779	7024	10457	836	3002	1968	2925	225	831	18279	24417	3031	6918	19992	29721	2347	8507
1999	3401	5230	359	1195	4313	6642	455	1520	8086	12478	881	2889	958	1477	102	339	28647	37098	3760	7621	22659	34941	2402	8013
2000	6913	9645	581	1501	6664	9322	534	1429	14895	20901	1288	3310	2291	3208	195	504	48055	61508	5250	9779	32314	45127	2731	7044
2000	1709	2339	151	359	2436	3338	215	513	7804	10704	714	1671	1818	2497	162	386	31037	39405	3536	6297	17331	23744	1559	3674
2001	1562	2518	118	360	3049	4901	231	699	7247	11840	581	1716	1896	3053	147	439	28083	40025	3313	6330	21764	35007	1668	5013
2002	1985	2454	109	355	3368	4162	185	603	12701	15693	703	2276	5282	6528	288	943	45027	52657	4206	8218	36597	45189	1988	6506
2003	1674	2749	91	402	3210	5273	177	774	11863	19544	660	2882	2704	4452	149	655	43027	60513	4206	8883	26116	42892	1429	6282
	1674		130	794	2171	7572	177	1142	4827	16992		2591	2062	7282			33349	81031		12691	13676	42892	1246	7163
2005		5264						1590	_		456				191	1107			4320	-	24532	4/3/6		
2006	3791	6397	498	1302	4627	7824	602		9554	16155	1271	3310	2986	5056	392	1032	46296	67532	8247	14807			3210	8400
2007	2063	4502	263	867	3047	6636	387	1275	4907	10706	636	2071	2442	5323	314	1027	29402	54734	4511	10780	17446	37934	2222	7293
2008	3285	5948	293	955	3971	7202	351	1057	9314	16832	841	2711	2178	3940	193	631	43277	66465	3580	9346	21887	39305	1915	6246
2009	2835	5834	198	1311	4193	8665	298	1957	6203	12763	442	2877	1450	2970	100	664	31106	50196	3526	10610	17820	36229	1200	8032
2010	5703	7334	496	1432	7062	9081	616	1774	6859	8842	604	1742	2606	3347	226	651	49703	58889	5478	10747	27468	35298	2358	6848
2011	4364	8077	433	2099	7477	13826	716	3566	5696	10554	564	2744	2074	3827	203	990	33849	62267	3160	15915	20249	37231	1953	9575
2012	5898	8720	471	1256	7488	11027	581	1566	5993	8819	468	1255	2348	3450	184	490	44778	65820	3395	9251	31467	46268	2451	6571
2013	3973	6228	254	2071	6681	10498	424	3502	6130	9621	398	3213	1701	2668	104	885	28314	44299	3301	13638	11746	18341	734	6050
2014	2222	3597	205	538	3340	5417	303	807	2701	4382	255	663	1374	2218	125	330	27090	43731	3558	8032	19410	31358	1782	4682
2015	3897	6679	375	1095	4182	7159	393	1164	4776	8201	465	1351	1622	2778	145	444	41399	70742	5248	12970	27736	47369	2666	7747
2016	2666	5928	301	957	3783	8189	405	1293	3906	8514	419	1347	1612	3499	163	543	22477	48428	3316	9211	17030	36504	1792	5712
201	7 1496	2974	147	397	2297	4559	220	603	2141	4232	211	564	1843	3656	164	471	20313	40217	2798	7209	17350	34228	1633	4488

Appendix 4.viii. Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run-reconstruction.

		Salmon Fis	shing Area	3		Salmon Fis	hing Area 4			Salmon Fis	hing Area 5			Salmon Fish	ing Area 6		5	Salmon Fis	shing Area	7		Salmon Fishi	ng Area 8	
		2SW		2SW	2	sw	25	W	25		28	w		w	2SW		28			SW	28	w	2SV	N
	Returns		Spawne	rs	Returns		Spawners		Returns		Spawners		Returns		Spawners	Ret	turns		Spawners		Returns	s	pawners	
Year		lin Ma		Min Max			Min	Max		Max		Max		Max	Min	Max	Min						Min	Max
Bugs labels	SFA3R2_L	[] SFA3R2_H[]	SFA3S2	_L[] SFA3S2_H[	SFA4R2_L[]	SFA4R2_H[]	SFA4S2_L[]	SFA4S2_H[]	SFA5R2_L[] S	SFA5R2_H[]	SFA5S2_L[]	SFA5S2_H[]	SFA6R2_L	SFA6R2_H[ S	SFA6S2_L SFA6	S2_H[ SF	A7R2_L	SFA7R2_H	SFA7S2_L	SFA7S2_H	SFA8R2_L	SFA8R2_HS	FA8S2_L 5	3FA8S2_H
197	0	15 1-	47	15 14	7 96	912	91	902	44	419	40	412	2	16	1	15	0	4	0	4	0	3	0	3
197	1	15 1-	40	14 13	7 75	711	69	700		316	29	308	1	10	1	10	1	8	1	8	0	5	0	3
197	2	10	94	10 9	4 68	648	66	643	37	356	35	352	2	22	2	22	1	11	1	11	1	5	1	5
197	3	23 2	23	23 22	3 132	1262	127	1252	42	397	40	395	5	47	5	47	3	25	3	25	2	18	1	17
197			29	32 12			198	810		252		247	12		12	46	5		-	20	2	-	2	8
197			08	52 20			263	1084	93	374		369	4		4	17	2	-		8	4		4	16
197			52	38 15			249	1035	76	305		292	10	40	10	39	1	5		5	3	1.0	3	13
197			93	48 19			156	841	145	582		573	13	51	11	47	2			10	3	10	3	10
197			60	15 6			123	592		226		220	6	23	5	22	2	- 8	_	8	1	3	1	3
197			56	39 15			121	548		102		98.	5	19	4	18	2			10			2	8
198			04	22 9			90	462		162		156	5	19/	3	16	2			6	_	,	2	7
198			18	104 41			564	2327	155	618		611	27	107	24	101	5					- 1	3	14
198			85	12 6			54	269	20	78		68	2	9	1	6	1	5	-	-		3	0	1
198			99	25 9			107	489	36	144		110	5	18	4	16	2	9		2	_	3	0	2
198			08	6 10			53	1064	15	291		291	2	34	1	33	1	10				-	0	9
198			37	7 13			67	1270	19	363		363	2	45	2	45	1	12		12	1	11	1	11
198			83	7 8			84	995	37	434		434	4	48	4	48	1	11		11	1	15	1	15
198			64	6 6			56	616		188		188	1	15	1	15	0	3		-		-	0	5
198			91	16 19			89	1073		454		454	4	47	4	47	1	14		14	_	9	1	9
198			92	9 9			46	473		177	17	177	3	31	3	31	1	8		8	-	12	1	12
199			84	24 18			78	607	33	259	33	259	5		5	36	1		-	5	-	9	1	9
199			50	19 15			72	558		233	-	233	3	21	3	21	1	4		4		1	0	1
199			19	42 81			141	2773		1018	52	1018	6	121	6	121	1	21		21		0	0	0
199			23	40 32			159	1245		456	58	454	9	70	9	70	2	1.0		13		12	1	12
199			57	45 45			99	1016		341	34	339	2	19	2	19	0	_		2		4	0	4
199			73	35 36			108	1139	40	406	39	403	1	14	1	14	1	9	_	8		7	1	
199			98	85 59			225	1594	77	533	76	530	3	20	3	19	1		_	5	-	7	1	
199			56	73 55			148	1132	40	306	40	305	1	11	1	11	0						0	3
199			79	116 97			310	2604	71	592	69	588	3	27	3	27	1	12		12	1	8	1	8
199			87	77 58			339	2602	58	440	4.1	438	2	17	2	17	0	3			1	6	1	- 6
200			22	87 52			168	1008		335		333	5		5	30	1	4	-	4	0	3	0	3
200			96	38 19			130	659	33	167		166	1	5	1	5	0		0		0	0	0	
200			87	29 18			94	604	11	69		69	1	4	1	4	0	0		0		1	0	
200			24	47 32			161	1116		99		99		7	1	7	0	2			0	1	- 0	
200			57	17 15			104	9/1	18	165		164		4	0	4	0	0				1	0	
200			49 01	20 24 53 29			116	1417 1193		206 240		205	0	4	0	4	0			-	-	5	- 0	
200							212	1193	43			237 297	0	2	0	2	- 0	0		- 0	1	3	1	
			56				201	1432		298			0	7		3	0	2		4		1 4	- 0	
200			16	45 31			204 150			260		260	1	17	1	17	0			0	1	5	- 1	
200			73	33 47			258	2114		377		377			3	17	0	0				-	0	
201			27	69 42 56 58			206	1609	76	469		468	3	20 32	3	20	0	4		4	U	_	- 0	10
201		56 58 71 4						2225	52 49	535 283		533 283	3	32 14	2	31 14	0	2				10 7	1	10
								1258			49		2		2		0	12		12		5	0	5
201		22 37 49 27		21 37			124 142	2165	43 30	731		728	2	35 10	2	35 10	0					2	0	5
201		49 27 73 45		48 266 72 455			288	799	61	167 380	29 60	165 377	4	22	3		0	3		3		8	- 0	
201		73 45 49 32		72 45: 46 32:			288 149	1805 1020	33	218		215	2		2	22 12	0			-		6	1	5
201		55 3°		53 31			115	675		218 99		215 98	0	3	0	3	0		0			3	0	3
201	/ _	JU 3	וט	31	118	080	175	6/5	17	99	17	98	U	3	U	3	U	1		U		3	U	اد

Appendix 4.viii. (Continued). Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run-reconstruction.

		Salmor	n Fishing	g Area 9			Salmon F	ishing Ar	ea 10		Saln	non Fishi	ng Area 11		S	almon Fis	hing Area 12	2	S	almon Fishi	ng Area 13		S	almon Fishing	Area 14A	
	2	2SW		2SW			SW		2SW		2SW		2SW		28	W	28		2SW		2SW		2SW		2SW	
	Returns			awners		Returns		Spawi		Return			pawners		Returns		Spawners		Returns		Spawners		Returns		awners	
Year	Mir		Max	Min	Max	Min				lax	Min	Max	Min	Max		Max		Max	Min	Max	Min	Max	Min	Max	Min	Max
Bugs labels																			SFA13R2_L[] SFA							
197			356	36	354	12	-	13		112	99	945	92	931		141		125	1300	3036	643	2050	36	514	13	468
197			305	30	301	18		74		171	80	763	74	750		85		82	1071	2850	653	2223	31	435	0	370
197			214	22	213	11		07		106	97	922	88	905		174		170	1243	3101	802	2439	20	280	8	257
197			406	41	402	35		36		327	96	913	92	906		121		116	1321	3120	764	2285	43	611	9	543
197			230	57	228	47		86		184	172	688	168	681		101		99	1165	2554	799	2005	90	361	79	338
197			352	87	351	20		80	19	78	212	846	208	838		96		92	1799	4454	1445	3923	51	203	42	185
197			261	64	258	29		17		115	167	669	163	660		48		47	1351	3293	1101	2917	144	575	134	555
197			136	33	134	35		38		137	129	516	127	513		70		61	1151	2159	731	1530	67	266	19	172
197			103	24	99	31		26		117	79	315	77	312		29	5	25	1886	3173	1544	2660	27	106	19	92
197			130	31	127	16		65	15	63	66	262	65	261		56	13	54	473	1117	428	1049	23	93	17	82
198			142	30	130	23		91	20	86	76	304	71	295		49	9	43	2094	3548	1697	2952	69	278	51	242
198			330	77	320	58		33		228	313	1252	309	1243		188	45	184	2059	4471	1794	4073	109	436	95	409
198			59	11	52	10		40	9	38	59	236	54	225		50	N	47	1377	2298	1139	1941	309	1238	299	1216
198			141	28	127	17		70	10	54	73	292	70	287		41	9	39	1786	3060	1542	2694	170	681	163	668
198			244	12	243	8		57		157	39	757	37	754		207	4	194	918	2226	795	2041	27	518	18	501
198			182	10	182	7		38		138	32	607	32	607	6	114	6	114	550	1505	540	1489	20	378	20	377
198			246	21	246	14		64		164	50	594	50	594	8	100		100	832	2190	805	2150	45	539	44	537
198			98	9	98	4		48	4	48	39	432	39	432	8	94	8	93	618	1813	605	1793	47	522	47	521
198			150	12	150	10		24		124	46	548	46	548	13	154	13	153	780	2350	764	2326	57	681	55	678
198			164	16	164	11		12		112	36	370	36	370	7	70		70	339	939	334	931	39	400	39	399
199	90 20	6	200	26	200	11		89	11	89	61	476	61	476	12	92		91	711	1851	698	1830	69	541	68	538
199	91 8	8	64	8	64	3		26	3	26	27	211	27	211	9	73		73	684	1460	676	1448	52	406	51	404
199		8	362	18	362	7		28		128	60	1192	60	1192	17	335		333	1235	5357	1197	5300	127	2505	123	2498
199	93 40	0	312	40	312	20	1	53	19	152	86	673	86	672	21	162		162	1047	2848	1018	2804	110	852	106	844
199	94 11	1	105	10	104	9		89	9	88	27	267	26	265	10	97	9	95	1390	3528	1283	3366	47	466	44	460
199	95 19	9	195	18	193	16	1	60	15	159	39	393	38	391	8	84	. 8	83	737	3058	643	2916	75	762	71	754
199	96 27	7	184	26	183	36	2	53	35	250	84	580	82	576	22	154	22	152	1391	4279	1280	4111	145	1004	141	996
199	97 23	3	172	22	171	32	2	40	31	238	104	787	103	784	28	215	28	214	1696	5113	1594	4960	158	1193	155	1185
199	98 28	8	234	27	233	47	3	90	46	389	51	422	50	421	14	118	13	116	1278	4248	1212	4151	144	1197	141	1191
199	99 22	2	169	22	167	28	2	15	27	213	53	405	53	404	6	48	6	48	1551	4643	1504	4573	149	1133	144	1122
200	00 36	6	212	35	210	35	2	06	32	200	79	466	77	463	12	71	12	71	2208	6029	2100	5867	168	995	164	986
200	01 7	7	34	7	33	10		49	9	48	31	156	31	155	7	37	7	36	697	2324	658	2248	69	346	67	342
200	02 5	5	34	5	33	10		66	10	65	25	160	25	160	6	41	6	41	642	2309	616	2260	74	472	72	466
200	03 5	5	33	5	33	8		57	8	56	31	213	30	212	13	89	12	88	822	3011	782	2934	89	612	85	605
200	)4 4	4	38	4	37	8		73	8	72	29	270	28	268	7	61	6	61	801	3255	758	3171	64	590	61	584
200	05 (	6	76	6	74	9	1	08	8	106	20	243	20	241	9	104	8	103	855	4629	804	4531	57	673	54	666
200	06 22	2	122	21	121	27	1	50	26	148	55	309	55	308	17	97	17	96	1581	5376	1534	5286	140	786	138	781
200	07 12	2	81	11	81	17	1	19	17	119	28	193	27	193	14	96	13	95	872	3912	839	3849	97	682	96	678
200	08 13	3	90	13	89	16	1	09	15	107	37	253	36	252	9	59	8	59	726	3451	666	3337	85	586	82	581
200	9	9	122	9	122	13	1	83	13	182	19	268	19	268	4	62	4	62	692	3858	656	3788	54	752	52	747
201	10 22	2	134	21	133	27	1	66	26	165	26	163	26	162	10	61	10	61	1078	3951	1019	3837	104	643	101	637
201	11 19	9	196	19	195	32	3:	35	31	332	25	256	24	255	9	92	9	92	144	1498	136	1480	86	895	84	890
201	12 20	0	117	20	117	26	1-	17	25	146	20	117	20	117	8	46	8	46	152	873	146	860	107	614	105	611
201	13 11	1	194	11	193	19	3:	28	18	326	18	300	17	299	5	83	4	82	854	3687	814	3621	33	567	32	563
201	14 9	9	51	9	50	14		77	13	75	11	62	11	62	6	31	5	31	799	2685	778	2651	80	442	77	435
201	15 17	7	103	16	102	18	1	10	17	108	20	127	20	126	7	43	6	41	1201	4245	1137	4142	117	726	115	720
201	16 14	4	90	13	89	18	1:	22	17	120	19	128	18	125	8	52	7	50	916	3613	829	3473	81	539	77	531
201	17 6	6	37	6	37	10		57	9	56	9	53	9	52	8	46	7	44	671	2623	639	2572	73	424	70	417

Appendix 4.ix. Input data for small salmon returns to Québec by category of data used in the run-reconstruction.

	Small returns								Small returns							
	Minimum								Maximum							
Year	C1	C2	C3	C4	C5	C6	FN Harvest	Other rivers	C1	C2	C3	C4	C5	C6	FN Harvest	Other rivers
Bugs labels	QCSmC1_L[] Q	CSmC2_L[] C	CSmC3_L[] QC	SmC4_L[] QC	CSmC5_L[]QC	SmC6_L[]	QCSmFn_L[]	QCSmO_L[]	QCSmC1_H[] Q	CSmC2_H[Q	CSmC3_H[	QCSmC4_H[	QCSmC5_H[	QCSmC6_H[	QCSmFn_H[]	QCSmO_H[]
1970	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
1971	0	0	0	0	0	0	0		0	0	0	0	0	0	0	0
1972		0	0	0	0	0	0		-	0	0	0	0	0	0	0
1973		0	0	0	0	0	0			0	0	0	0	0		0
1974	0	0	0	0	0	0	0		0	0	0	0	0	0	0	0
1975		0	0	0	0	0	0			0	0	0	0	0	0	0
1976		0	0	0	0	0	0			0	0	0	0	0		0
1977		0	0	0	0	0	0			0	0	0	0	0		0
1978		0	0	0	0	0	0			0	0	0	0	0		0
1979		0	0	0	0	0	0			0	0	0	0	0		0
1980		0	0	0	0	0	0		-	0	. 0	0	0	0	0	0
1981		0	0	0	0	0	0		_	0	0	0	0	0	0	
1982		0	0	0	0	0	0			0	0	0	0	0	0	0
1983		0	0	0	0	0	0			0	0	0	0	0	0	0
1984		5434	2955	460	1670	5160	267		4085	5639	6053	792	2784	8599	445	52
1985		2271	1767	210	5449	4384	267			2336	3586	352	9224	7307	445	67
1986		5193	2396	63	6719	5133	267	_		5321	4895	107	11198	8555	445	129
1987		4775	3852	327	8396	5501	267	71		4910	7875	546	13993	9168	445	118
1988		5968	4404	468	8440	6423	267	85		6133	8962	780	14067	10705	445	142
1989	10121	4743	2924	301	6744	5622	267	68	10910	4878	5940	503	11240	9369	445	113
1990		7332	4377	694	7096	2976	377			7511	8917	1158	11826	4960	628	129
1991	9554	5851	3776	349	5009	2001	256			5987	7679	584	8348	3336	426	95
1992	9188	6928	4567	428	5131	3462	243	_		7144	9297	715	8552	5770	405	117
1993	8143	6325	3973	1029	4315	1447	525			6517	8075	1717	7192	2412	875	92
1994		5928	3840	1051	4011	437	408			6129	7828	1753	6686	729	681	50
1995		3439	2697	1017	3853	434	184			3527	5471	1696	6422	723	306	50
1996	15010	1809	3600	477	4666	500	120			1923	7370	797	7816	833	200	8
1997		201	3457	292	3529	462	58			242	7049	487	5882	770	97	938
1998		1183	3578	328	5121	1127	58			1406	7347	555	8536	1878	97	0
1999		708	3194	1868	5401	1429	0			741	6536	3098	9002	2382	0	0
2000		429	1116	602	7399	633	0			458	2284	1004	14050	1055	0	0
2001		185	2632	266	3225	728	0			228	5392	443	5374	1213	0	0
2002		31	3189	689	4333	1448	0			36	6530	1149	7222	2414	0	0
2003		0	3203	721	3566	1512	0			0	6538	1201	5944	2520	0	0
2004		107	6526	284	4889	1639	0			127	13104	474	8149	2731	0	0
2005		0	3689	794	3353	1508	0			0	7485	1323	5588	2513	0	0
2006		0	3736	1800	2944	1455	0			0	7584	2999	4907	2426	0	0
2007		0	3758	1710	1830	1024	0			0	7631	2850	3051	1707	0	0
2008		0	5542	2266	3144	1401	0			0	11261	3776	5240	2336	0	0
2009		0	3601	903	1907	1056	0			0	7306	1505	3178	1759	0	0
2010		0	4801	993	1675	1081	0			0	9746	1655	2792	1802	0	0
2011		0	5120	1365	4441	1694	0			0	10386	2276	7402	2824	0	0
2012		0	3615	584	3550	1228	0			0	7332	973	5916	2047	0	0
2013	7219	88	3185	411	2466	1401	0	0	7574	104	6461	685	4111	2335	0	0
2014	9193	0	3945	676	2412	2341	0		9647	0	8003	1127	4020	3901	0	0
2015	17189	0	5843	1253	3772	1827	0	0	18048	0	11853	2088	6287	3046	0	0
2016	13183	0	5977	1049	3202	2756	0	0	13850	0	12124	1749	5336	4593	0	0
2017	8111	0	4562	503	3673	1787	0	0	8528	0	9254	839	6121	2978	0	0

Appendix 4.ix. (Continued). Input data for large salmon returns to Québec by category of data used in the run-reconstruction.

	Large return	าร							Large retur	ns						
	Minimum								Maximum							
Year	C1		C3		C5			Other rivers		C2	C3	C4	C5	C6		Other rivers
		QCLgC2_L[]							-		] QCLgC3_H[]	QCLgC4_H[]	QCLgC5_H[]	QCLgC6_H[]	QCLgFn_H[]	QCLgO_H[]
1970		-	0								0 (	, .	_			
1971			0				) (				0 (					-
1972			0				) (				0 (					-
1973			0	-	(						0 (	, ,	-	-		
1974			0		(		) (				0 (	, ,			-	
1975			0		(		) (				0 (	, ,				
1976		-	0	-	(				) (		0 (	, ,	-			
1977			0		(				) (		0 (	, .			-	
1978			0		(				,		0 0	, ,		-		U
1979		-	0								0 (	, ,			-	
1980		-	0		(		) (				0 (	,				U
1981	0	-	0	-	(						0 (	, ,	-	-		
1982			0		(		) (				0 (					
1983			0		(		,					0				-
1984			2922		3712											
1985			3836		9215											
1986			6152		5877											
1987	17574	6715	5178		6335				_				10558	5019		
1988			7540		6789				22979							
1989	20278	8503	5530	278	5718	456	7 329	106	21906	873	6 11252	2 465	9531	7611	548	176
1990	17098	10803	8164		5179								8631			
1991			7183		3856			2 101	20443							
1992	18392	7360	7930	372	2687			1 76	19578	756	0 16149	622	4478	2505	769	127
1993			2866		2649											
1994	16538	9172	2644	506	2853	14			17594	1024	1 541	845	4755	242	712	
1995	21658	9598	1926	813	4390	15-	1 246	31	22968	1093	6 391	1358	7317	256	410	
1996			3843		2486				24117							
1997	18106	4221	2816	333	2865				19154	1 515	4 5768	553	4775	229		
1998	13180	4927	2861	347	2790			3 (	13891	596	2 5907	592	4649	485	80	0
1999	16912	842	2554	3661	3870			) (		99	5 5232	2 6103	6450	838	0	0
2000	14568	619	3901	560	6420				15300	66	9 7947	933	10700	949	0	0
2001	17837	633	5320	241	3988	550	3 (	) (	18889	87	9 10914	402	6647	926	0	0
2002			4515		2103			) (			9 927					
2003			5787								0 11779					
2004	18369	107	4870					) (		12	6 9170	595	7387			0
2005			3204						20066		0 651					
2006			3387	901	3945			) (		)	0 6904				0	0
2007	14832	0	3638	1301	3171	309	5 (	) (	15604	1	0 7406	2168	5285	508	0	0
2008	15216	0	5187	1328	5423	390	) (	) (	16002	2	0 1059	2213	9038	649	0	0
2009	18479	0	3727	950	4556	27	5 (	) (	19412	2	0 7589	1584	7594	458	0	0
2010	21375	0	4488	1047	3656	338	3 (	) (	22454	1	0 915	7 1744	6093	564	0	0
2011	26977	0	4697	1571	6007	483	3 (	) (	28373	3	0 9529	2619	10011	805	0	0
2012	17918	0	3665	904	4488	31:	3 (	) (	18837	7	0 7434	1 1507	7481	522	0	0
2013	22026	205	4171	989	3938	339	9 (	) (	23135	5 24	2 846	1 1648	6563	565	0	0
2014	10954	0	2400	695	1593		5 (	) (			0 4869	1159	2655			0
2015	18449	0	3995	1641	3943			) (	19388	3	0 8104	1 2735	6572			0
2016			4244		5022			) (	18785		0 8608					0
2017			4639		4926				20023		0 941					

Appendix 4.ix. (Continued). Input data for small salmon spawners to Québec by category of data used in the run-reconstruction.

	Small spawn	ers					Small spawn	ers				
	Minimum						Maximum					
	C1	C2	C3	C4	C5	C6	C1	C2	C3	C4	C5	C6
Bugs labels	QCSSmC1_L[	QCSSmC2_L	QCSSmC3_L	QCSSmC4_L[	QCSSmC5_L	QCSSmC6_L	QCSSmC1_H	QCSSmC2_H	QCSSmC3_H	QCSSmC4_H	QCSSmC5_H	QCSSmC6_H
1970	0	0	0	0	0	0	C	0	0	C	0	0
1971	0	0	0	0	0	0	C	0	0	C	0	0
1972	0	0	0	0	0	0	C	0	0	C	0	0
1973	0	0	0	0	0	0	C	0	0	C	0	0
1974	0	0	0	0	0	0	C	0	0	0	0	0
1975	0	0	0	0	0	0	C	0	0	0	0	0
1976	0	0	0	0	0	0	C	0	0	0	0	0
1977	0	0	0	0	0	0	0	0	0	0	0	0
1978	0	0	0	0	0	0	0	0	0	0	0	0
1979	0	0	0	0	0	0	0	0	0	0	0	0
1980	0	0	0	0	0	0	C	0	0	0	0	0
1981	0	0	0	0	0	0	0	0	0	C	0	0
1982	0	0	0	0	0	0	9	0	0	0	0	0
1983	0	0	0	0	0	0	0	0	0	0	0	0
1984	3061	4342	1915	415	1264	5160	3316	4547	5013	747	2378	8599
1985	3960	1622	1025	209	4241	4384	4563	1687	2844	351	8016	7307
1986	6337	3827	1499	63	5151	5133	7160	3955	3998	107	9630	8555
1987	7493	3489	2365	291	6411	5501	8319	3624	6388	510	12008	9168
1988	8173	4188	2738	419	6432	6423	9227	4353	7296	731	12059	10705
1989	7779	3810	1878	273	5149	5622	8568	3945	4894	475	9645	9369
1990	8735	5757	2822	604	5437	2976	9768	5936	7362	1068	10167	4960
1991	7247	4551	2465	316	3827	2001	7942	4687	6368	551	7166	3336
1992	5989	4841	2937	370	3957	3462	6648	5057	7667	657	7378	5770
1993	4852	4311	2524	747	3339	1447	5592	4503	6626	1435	6216	2412
1994	5506	3996		894	3089	437	6241		6489			729
1995	5348	2835			2956		5943			1556		723
1996	10636	1330			3678		11748					833
1997	8238	142	2250	266	3074	462	8836	178	5842	461	5426	770
1998	7734	995		. 289	4229	1124	8298	1218				1875
1999	8155	509	2495	1653	4581	1426	8834	542	5837	2883	8182	2379
2000	8291	372			5900	583	9040		1861	921		1005
2001	5329	143	1870		2579	658	5867	186	4140	440	4729	1137
2002	9296			658	3405						6294	2414
2003	8180				2826		8721				5204	2517
2004	9030		$\overline{}$		3962		9460					
2005	6339				2709		6756			1245		2511
2006	8628			1691	2372	1455	9235	5 0		2890	4335	2426
2007	5768				1501	1024	6217				2722	
2008	10562				2522		11467			3266		
2009	6293				1633							1759
2010	8860	0	_		1311	1080	9509			1576		
2010	12143				3674							2818
2012	6620	0			2924	1225					5290	2044
2012	4959	88			2924	1401	5314					
2013	6579	00			2127	2175						3735
2014	11926	0			2935		12785				5450	3039
2015	9541	0			2484	2591	10208					4428
2016	5810			441	2806		6227					

Appendix 4.ix. (Continued). Input data for large salmon spawners to Québec by category of data used in the run-reconstruction.

	Large spaw	ners					Large spaw	ners				
	Minimum						Maximum					
Year	C1	C2	C3	C4	C5	C6	C1	C2	C3	C4	C5	C6
Bugs labels	QCSLgC1_L[	QCSLgC2_L[	QCSLgC3_L[	QCSLgC4_L[	QCSLgC5_L[	QCSLgC6_L[	QCSLgC1_H	QCSLgC2_H	QCSLgC3_H	QCSLgC4_H	QCSLgC5_H	QCSLgC6_H
1970	0	0	0	0	0	0	C	0	0	0	0	0
1971	0	0	0	0	0	0	C	0	0	0	0	0
1972	0	0	0	0	0	0	C	0	0	0	0	0
1973	0	0	0	0	0	0	C	0	0	0	0	0
1974	0	0	0	0	0	0	C		0	0	0	0
1975	0	0	0	0	0	0	C		0	0	0	0
1976	0	0	0	0	0	0	C	d	0	0	0	0
1977	0	0	0	0	0	0	0	0	0	0	0	0
1978	0	0	0	0	0	0	C	O	0	0	0	0
1979	0	0	0	0	0	0			0	0	0	0
1980	0	0	0	0	0	0		0	0	0	0	0
1981	0	0	0	0	0	0		0	0	0	0	0
1982	0	0	0	0	0	0		0	0	0	0	0
1983	0	0	0	0	0	0		0	0	0	0	0
1984	10421	7648	1861	2357	2815	5071	11933	7935	4974	5427	5290	8452
1985	9985	4991	2125	340	7214	3351	11581	5244	6098	572	13826	5586
1986	13659	5804	3695	35	4498	4971	15672	6045	10139	61	8416	8284
1987	13432	4791	3025	246	4830	3012	14875	4965	8422	431	9053	5019
1988	15535	4258	4381	312	5172	4781	17069	4444	12177	542	9698	7969
1989	14645	6742	3239	253	4375	4567	16273	6975	8961	440	8188	7611
1990	12398	8463	4557	1228	3950	2424	13522	8701	13006	2139	7402	4040
1991	14061	5019	3970	596	2940	357	15392	5223	11389	1061	5511	595
1992	12850	4819	4492	325	2044	1503	14036	5019	12711	575	3835	2505
1993	9848	6936	1809	282	2038	333	10724	8266	4792	533	3803	555
1994	10468	5920	1693	448	2173	145	11524	6989	4460	787	4075	242
1995	16562	8323	1321	781	3367	154	17872	9661	3310	1326	6294	256
1996	16431	4417	2389	394	1924	135	17869	5536	6390	781	3593	225
1997	13433	3393	1744	308	2237	138	14481	4326	4696	528	4147	229
1998	10402	4429	1849	302	2213	290	11113	5464	4895	547	4073	484
1999	14169	747	1962	3100	2956	491	14957	900	4640	5542	5536	837
2000	11937	570	3322	491	5096	363	12669	620	7368	864	9376	749
2001	14527	505	4281	239	2980	348	15579	751	8986	400	5639	717
2002	10843	8	4071	313	1500	344	11509	9	8833	539	2902	574
2003	18832	0	5164	267	3763	383	19872	: C	11156	447	7022	640
2004	15558	107	4231	355	3268	401	16232	126	8531	593	6223	668
2005	16485	0	2901	719	3556	351	17397	0	6212	1208	6766	585
2006	14977	0	3055	872	2863	403	15773	0	6572	1473	5493	672
2007	12470	0	3203	1287	2444	303	13242	. 0	6971	2154	4558	506
2008	13725	0	4676	1266	4296	390	14511	0	10084	2151	7911	649
2009	16489	0	3188	849	3588	275	17422	. 0	7050	1483	6626	458
2010	19170	0	3926	1023	3017	338	20249	0	8595	1720		564
2011	24130	0	4180	1497	4579	479	25526	0	9012	2545	8583	801
2012				868						1471	6678	522
2013		205		920						1579		564
2014	10089									1128		1430
2015	17003									2673		956
2016										3162		1691
2017		0								1753		853

Appendix 4.ix. (Continued). Year specific returns and spawner data for Québec for years when category splits are not available (1970 to 1983) used in the run-reconstruction.

	Small salm	on	Large salmo	on	Small	returns		Large returns		Small spa	wne	ers	Large spawne	ers
Year	FN	Commercial	FN	Commercial		Min	Max	Min	Max		Min	Max	Min	Max
Bugs labels	QCFnSm[]	QCCmSm[]	QCFnLg[]	QCCmLg[]	QCSm	L[]	QCSm_H[]	QCLg_L[]	QCLg_H[]	QCSSm_L	-[]	QCSSm_H[]	QCSLg_L[]	QCSLg_H[]
1970	C	0	0	0		18904	28356	82680	124020	11	1045	16568	31292	46937
1971	C	0	0	0		14969	22453	47354	71031	9	9338	14007	16194	24292
1972	C	0	0	0		12470	18704	61773	92660	8	3213	12320	31727	47590
1973	C	0	0	0		16585	24877	68171	102256	10	987	16480	32279	48419
1974	C	0	0	0		16791	25186	91455	137182	10	0067	15100	39256	58884
1975	C	0	0	0		18071	27106	77664	116497	11	1606	17409	32627	48940
1976		0	0	0		19959	29938	77212	115818		2979	19469	31032	46548
1977	C	0	0	0		18190	27285	91017			2004	18006	44660	66990
1978	C	0	0	0		16971	25456	81953	122930		1447	17170	40944	61416
1979	C	0	0	0		21683	32524	45197	67796		5863	23795	17543	26315
1980		0	0	0		29791	44686	107461			0817	31226	48758	73137
1981	C	0	0	0		41667	62501	84428	126642		0952	46428	35798	53697
1982						23699	35549	74870			877	25316	36290	54435
1983			0	0		17987	26981	61488	92232		2030	18045	23710	35565
1984	357		4530	13053		0			0		0	0	0	0
1985			3623	16619		0		0	0		0	0	0	0
1986			4519			0		.0			0	0	0	0
1987	366		4466	22745		0		0	0		0	0	0	0
1988	397		4747	19750		0		0	0		0	0	0	0
1989			2905	18175		0		0	-		0	0	0	0
1990	108		2900	16092		0					0	0	0	0
1991	265		4335	16372		0	0				0	0	0	0
1991			4550	15851		0					0	0	0	0
1992	7		3976	11242		0					0	0	0	0
1993			4496	10424		0					0	0	0	0
1994	353		6194	10424		0					0	0	0	0
1995						0					0	0	0	0
1996	72 35		6113 4875	7454 7202			0				0	0	0	0
						0					0	0		
1998	35		4875			0							0	
1999			3683								0	0		0
2000			3818		_	0					0	0	0	0
2001	770			_		0					0	0	0	0
2002				0		0					0	0	0	
2003	972					0	0				0	0	0	0
2004	1158					0	0				0	0	0	0
2005	909	_				0					0	0	0	
2006	1117		3512			0					0	0	0	
2007	869		2932			0					0	0	0	0
2008	1171			0		0	0				0	0	0	0
2009	1141			0		0	0				0	0	0	0
2010				0		0					0	0	0	0
2011	1205			0		0					0	0	0	0
2012				0		0	0	0			0	0	0	0
2013	1177			0		0	0				0	0	0	0
2014	1240			0		0	0				0	0	0	0
2015	1246			0		0	0				0	0	0	0
2016	1277			0		0	0				0	0	0	0
2017	1191	0	2677	0		0	0	0	0		0	0	0	0

Appendix 4.x. Input data for 2SW, large, and small salmon returns to Salmon Fishing Areas 15 to 18 for Canada used in the run-reconstruction.

Year	SF15R2 L	SF15R2 H	SF16R2 L	SE16R2 H	SF17R2 L	SF17R2 H	SF18R2 L	SF18R2_H	SF15Lg L	SF15Lg H	SF16Lg L	SF16Lg H	SF17Lg L	SF17Lg H	SF18Lg L			SF15Sm_ H		SF16Sm_ H	SF17Sm_ S		SF18Sm_	SF18Sm_ H
1970			42901			60	4744			16270			31				2834	6279	47779	67697	0	0	264	1073
1971	3587	4616	26038	30669	29	29	1891	2782	5518	7102	28365	33409	29	29	2456	3198	2113	4681	38388	54120	0	0	65	265
1972	4980	9756	29092	43510	402	402	4693	6024	8441	16536	30146	45087	402	402	6095	6924	2185	4699	48886	69270	0	0	131	530
1973	6211	12009	26599	40492	206	206	4140	5481	8393	16229	27771	42276	206	206	5376	6299	3010	6668	47190	66835	5	9	516	2095
1974	7264	14570	39270	60090	386	386	5481	6928	9950	19959	43249	66179	386	386	7119	7963	2226	4895	78091	110470	0	0	187	757
1975	4353	7922	25889	39325	345	345	3452	4340	5510	10028	29826	45305	345	345	4483	4989	2393	5298	69993	98443	0	0	112	454
1976	7293	14416	20448	30758	575	578	2755	3674	9596	18969	23943	36016	575	578	3578	4223	8667	14696	96504	136107	14	28	299	1212
1977	9174	18077	49881	73330	606	606	3985	5463	11053	21779	52673	77434	606	606			6085	12084	30621	42689	0	0	215	871
1978			19504				4585			14332	22653	30245		0			4350			39927	0	0		316
1979			6501			463	1290			4971		13507		_			4378			70714	2	5		7536
1980	7102				2		3732					51008		_		_	7994			58839	12	23		2108
1981	4572		11144			77	2490			15652		28887	40				9380				259	498		11348
1982	4314		21442				4135					51475					6541		78072		175	336		8722
1983	3453		16349				3733					35304	_	32		-	2723				17	32		
1984	3329		12216				2391					33321					12003				17	32		1867
1985	4805		14614			15	921		7627	15080	16798	43247	8				7003				113	217		3167
1986	7831		21617				2274			20267	25342	65228	5				10813				566	1088	965	3854
1987	4836		12524			128	2446			14255	15734	40483		_			9630		74485		1141	2194	1541	8586
1988	7152		14384				2365			19441	17627	45267	96	_			13168				1542	2963	1297	7353
1989 1990	4390		9113 14269			287	1970 1778		7701 6362	14898 12307	13955 23164	358 <b>12</b> 59479					6357				400 1842	770 3539		4843
	4326					545			4773					545			7880						921	5335
1991 1992	2387 4002	4668 7787	14685 21381			361 183	2181 2229			9335 14420	24273 34573	49686					4441 8853			92327 204708	1576 1873	3028 3599	1089 936	6266 5335
1992	1395		15579			43	1266			6711	22602	87407	-				5783		70090		1277	2454	1085	5900
1994	3960		13652		169	310	1866		6600								9136				210	385		3640
1995	2713		25593			576	1395			8199	30324	_					2902			63453	658	987	509	3006
1996	3917		11126			591	2931				16317	_					6034			45995	710	1065		13433
1997	2488		8545			581	3174			$\overline{}$		24521	387	581			5797			24122	517	776		3183
1998	1687	3260	6476			577	1921			_							6288			31988	508	762		3631
1999	1780		7369			575	1457			5269	_						4936				413	620		4011
2000			7916				1336			6785	16701	23743					7459				395	593		
2001	3779						1558		$\overline{}$								4947				415	622		4610
2002	2335		5686			557	1134		3592		11322	16883	372				11719				390	585	757	4821
2003	3947	7778	11231			557	2203		6072	11966	19192	27338					3119				515	773		4498
2004	3005	5886	10985	16899	367	550	2441	11019	4623	9055	19272	29647	367	550	3170	12665	12091	26783	45176	61718	330	495	988	6339
2005	3422	6725	11726	18853	373	560	2039		5265	10346	17634	28350	373	560	2649	10331	4117	9116	29727	45609	343	514	723	4472
2006	2551		10234	15570	392	587	1955		3924	7651	19681	29943	392	587	2539	10038	8724	19322	31229	51615	331	497	786	5003
2007	4267	8422	9827	14438	412	618	1345	6045	6565	12957	16655	24471	412	618	1747	6949	4259	9430	23584	41581	275	413	650	4123
2008	2848	5572	6156	10563	429	644	2031	9294	4382	8572	10993	18862	429	644	2638	10683	13601	30129	26405	44198	298	447	1127	7400
2009	3948	7781	11010	15595	402	602	1522	7153	6074	11970	17759	25154	402	602	1977	8222	5169	11445	11377	22184	233	350	267	1992
2010	2978	5831	8076	11204	439	658	2051	9512	4581	8972	15836	21968	439	658	2664	10933	8187	18132	49363	66530	258	387	783	5330
2011	7265	14445	22193	41736	653	980	3479	15787	11177	22223	26109	49101	653	980	4518	18146	10234	22668	42560	67042	291	436	1065	6965
2012	3230	6338	8402	14247	653	980	835	4083	4969	9750	12002	20353	653	980	1084	4693	4350	9631	7650	13982	291	436	184	1501
2013	5324	10544	7349	15519	719	1077	1848	8498	8190	16222	10651	22491	719	1077	2400	9768	4661	10320	10374	19192	274	410	394	2743
2014		5282	5798	11586	491	735	1250	5924	4160	8126		17190	491	735	1623	6809	3441	7616	6610		222	332	253	1915
2015			9723			846	1752										6135				235	349		4731
2016			10800			688	1640			9201				688			3921			23345	180	268		2544
2017	2416	4703	8713	15697	459	687	962	4910	3717	7235	12560	22628	459	687	1249	5644	3570	7902	11934	18851	203	303	285	2389

Appendix 4.x. (continued). Input data for 2SW, large, and small salmon spawners to Salmon Fishing Areas 15 to 18 for Canada used in the run-reconstruction.

									SF15SLg_	SF15SLg_	SF16SLg_	SF16SLg_	SF17SLg_	SF17SLg_	SF18SLg_	SF18SLg_	SF15SSm	SF15SSm	SF16SSm	SF16SSm	SF17SSm	SF17SSm	SF18SSr	m SF1	i8SSm
Year	SF15S2_L	SF15S2_H	SF16S2_L	SF16S2_H	SF17S2_L	SF17S2_H	SF18S2_L	SF18S2_H I	L	Н	L	Н	L	Н	L	Н	_L	_H	_L	_H	_L	_H	_L	_H	
1970	1156	3252	5346	8242	18	47	304	1587	1779	5003	5790	8926	18	47	395	1824	1417				0	(	) 1	L67	842
1971	510		6724			-			785		7324	12369	0			_	1056				0		-	41	208
1972	2367		17031						4011		17648		0		250						0		-	82	416
1973	2873		19277						3883		20126	34632	0	-							3			325	1645
1974	3620		31192						4960		34352	57282	0	-							0			18	595
1975	1769		18536						2239		21355	36834	0	_					50438		0		_	71	357
1976	3530		11842		_				4644		13867	25940	1				2480				8	22		188	951
1977	4412		30623			-			5315		32337	57097	0	_	257	_	2467				0			L35	684
1978	2622		6998						3496		8128	15720	0		290	_			14689		0		_	49	248
1979	527	1482	3000						1033		4355	8426	3		149	_					1		4 11		5915
1980	3440		17667						4248		18597	32590	1	4			2996			44943	7			327	1655
1981	1380		2392		_				2935		3586	15765	36	73	255	1181	3183		35423		151			_	8908
1982	991		8418						1679		10405	35376	8	23					51324		102			_	6847
1983	906		5516						1535		6852	21846	15			-	820				10		_	133	674
1984	2656		11650						3362		12341	32721	13	4	337						10			.77	1200
1985	4514		14019						7164		16114	42563		_							66			45	1788
1986	7279		20606		_				9577		24157	64044	5				5589		71918		330			63	1729
1987	4122		11414						6441	13099	14340	39088		128							665			122	4394
1988	6582		13801						9141		16913		96								899			260	3467
1989	3944		8466		_				6919		12965	34822	_		2471		3191				233		_	.74	2368
1990	3886	7903	13669						5715		22190		284				3996				1074			167	2510
1991	2193		14200		_				4386	_	23472		188								919			199	2933
1992	3639	7400	20770						6738		33583		95								1092			131	2320
1993	1239								3099		22109		22								745			200	2583
1994	3639	7401	13418						6065		17787	32682	166								118			135	1798
1995	2519								3873		30007	43778	380				1451				250			14	1527
1996	3688		10743					_	5674		15755		388				3017				258			315	8090
1997	2316		8106						3563		13955	23677	385			_					256			160	1841
1998	1512		6280					$\overline{}$	2326	-	15177	23666	382				3144				255		_	183	2113
1999	1581		6858						2433		14092	20732	379				2465				253			291	2669
2000	2057	4184	7614						3165	6437	16065	23043	376								252			254	2393
2001	3521	7161	14081						5417		21809	28427	374			_					250			317	3040
2002	2120		5446						3261		10843	16354	371				5857				249		_	340	3256
2003	3683	7491	10830						5666		18508	26580	368								248			327	3088
2004	2770		10529				_		4261	8666	18473	28755	365								246			891	4106
2005	3175		11167					_	4884		16792	27413	371				2056		17965		246			242	2669
2006	2329		9787						3583		18822	28991	390				4359				247		_	276	3092
2007	3994		9335						6145		15823	23568	409				2127		14802		248			260	2661
2008	2618		5719						4028		10212	18010	429								249			273	4200
2009	3684		10488		_				5668		16916	24245	401						6257		233			69	1251
2010 2011	2743		7605 21413						4221 10619	8584 21597	14912	20989 47977	438						31374 27398		256			192	3118 4343
2011	6902								4597		25192		652			_	5114				290 290			865	4343 1156
	2988	6077	8013								11447	19722	652				2172							92	
2013	5019		6766						7721		9805	21539	717								272		_	152	1837
2014	2478		5709						3812		8470	16981	489				1717				220			139	1488
2015 2016	3165 2815		9549 10539						4869 4331		13944 14642	23414 31249	562 458				5945 3797	13336 8518			229 176			120 213	3730 2066
2016	2815		8509		_				3382		14642		458 457				3/9/	7754	12123		201	30:	_	162	1928
2017	2198	44/1	8509	15431	. 45/	685	944	485/	3382	68/9	12266	22243	45/	685	1226	5583	345/	//54	1109/	1/952	201	30	4 T	.02	1978

Appendix 4.xi. Input data for 2SW, large, small salmon returns to Salmon Fishing Areas 19 to 21 for Canada used in the run-reconstruction.

	Returns of 2S	W			Returns o	of large sal	mon		Returns of sn	nall salmon		
	SFA 19-21		SFA 23		SFA 19-21	1	SFA 23		SFA 19-21		SFA 23	
Year of return												
to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
	SF19_21R2_L	SF19_21R2_H			SF19_21L	SF19_21L	SF23Lg_L	SF23Lg_	SF19_21Sm_	SF19_21Sm	SF23Sm_L[	SF23Sm_H[
Ninbugs labels	[]	[]	SF23R2_L[]	SF23R2_H[]	g_L[]	g_H[]		H[]	L[]	_H[]	]	]
1970	5600	7447	8540	12674	7273	9671	9691	13945	16177	24106	5306	7521
1971	4120	5215	7155	10536	5350	6773	8056	11573	11911	18004	3248	4541
1972	5744	6993	7869	11368	7460	9082	8890	12536	11587	17992	1831	2506
1973	6922	8659	4205	6036	8049	10069	4760	6638	14169	22159	5474	7012
1974	13138	15363	10755	14988	13138	15363	12187	16444	25032	39058	10195	12901
1975	12261	13797	13107	18578	12261	13797	14829	20351	10860	15753	18022	23101
1976	8607	10104	14274	20281	8873	10416	16128	22175	21071	33009	22835	28864
1977	10872	12851	16869	23995	14119	16690	19165	26183	24599	37314	13738	16671
1978	8272	9779	8225	11294	10471	12378	9335	12342	7621	10023	6271	7695
1979	3781	4879	5165	7207	5180	6684	5856	7903	24298	37514	15356	20517
1980	14094		19056	26865	16388		21464		34377			
1981	8662		11026	15267	11706		12481					
1982	4458		9782	13871	9485		11147					15636
1983	4134		9662	13836	6562	_	10908		9313			
1984	1758		15706	22627	2408		17706					
1985	6894		16541	23828	8512		18582				13056	
1986	6755		9891	14261	10722				24369			
1987	3748		6922	10043	5950		7865					
1988	4393		4716	6697	7321		5360					23084
1989	4808		6560	9437	6969		7393		25602		17579	
1990	3591	6320	5486	7918			6235					
1991	2960		7337	10563	4112		8312					17685
1992	2633		6878				7749					
1993	2542		4345				5260				7610	
1994	1360		3084	3495			3659					
1995	2253		3439	3998			3728					
1996	3000		4729	5397	3571		5535					15256
1997	1163		2769				3210					
1998	924		1372	1642			2032		8833			
1999	1419		2375		1709		2734		3971			
2000	1078		988 1938	1206	1315		1189					
2001	1822 382			2279	1980		2113		2326			
2002			483	548	749		639					3991
2003	1854		1056	1198	1952		1128		2844			
2004	1028		1335	1605			1402					
2005	662		809	1012			890		2870		3597	
2006	1263		922	1171	1559		997					
2007	603		616	736			689		4198			
2008	1793		812	1042			858				5924	
2009	827		1485	1886	1034		1678					
2010	934		829	992			1117					
2011	1489		2486	3259	1504		2598		3615			
2012	661	903	268	331	785		335					
2013	2075		420	543			503					
2014	415		172		415		230					
2015	338		248	314			257					
2016	883		415	515			457	568	284	383		
2017	493	678	487	620	513	705	517	660	2356	3181	1003	1293

Appendix 4.xi. (continued). Input data for 2SW, large, small salmon spawners to Salmon Fishing Areas 19 to 21 for Canada used in the run-reconstruction.

	Spawners					s of large s				s of small s		
	SFA 19-21		SFA 23		SFA 19-21		SFA 23		SFA 19-21	l .	SFA 23	
Year of return												
to rivers	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max	Min	Max
	SF19_21	_	_	SF23S2_	SF19_21	_		SF23SLg		_		SF23SS
Winbugs labels	S2_L[]				SLg_L[]				SSm_L[]			m_H[]
1970	2388					5499			9429			6101
1971	1418	2513	3612	6576	1841	3264	3888	7405	7246	13339	1216	2509
1972	1616	2865	6472	9806	2099	3721	7246	10892	7616	14021	0	1
1973	2246	3984	2752	4412	2612	4632	3050	4928	9502	17492	4037	5575
1974	2878	5103	8123	12046	2878	5103	9090	13347	16680	30706	8071	10777
1975	1987	3523	10987	16209	1987	3523	12335	17857	5819	10712	15363	20442
1976	1935	3432	10071	15583	1995	3538	11183	17230	14196	26134	17572	23601
1977	2559	4539	12013	18568	3324	5895	13452	20470	15120	27835	9196	12129
1978	1948	3455	5346	8076	2466	4373	5948	8955	2857	5259	4256	5680
1979	1419	2517	3772	5650	1944	3448	4217	6264	15716	28932	11640	16801
1980	4170	7394			4849	8598	13190	21206			19597	25941
1981	3631	6439	3642	7014	4907	8702	3794	8056	21096	38837	7805	12782
1982	1158											10357
1983	1579				2506							8454
1984	1416					_			13658			16412
1985	6761								18024		8164	13036
1986	6624				_						10725	16634
1987	3676				5835				20213			14561
1988	4322								18125			19764
1989	4735											20066
1990	3530								22080			15381
1991	2912											14139
1992	2588											14326
1993	2493		3291									6868
1994	1339									4900		5738
1995	2218											9218
1996	2946								10909			13892
1997	1140		_									4433
	915		1068									10348
1998	1409		1934						3895		8775 5196	6048
1999	1072				1697							
2000												5087
2001	1812	_										2530
2002	378											3689
2003	1834					2661						2469
2004	1017		1238									4039
2005	646											4450
2006	1248					2123						4501
2007	587											2937
2008	1778								7252		5729	7467
2009	811								2051	2773		1864
2010	910											11901
2011	1467									4864		5867
2012	641					1048			343		167	208
2013	2057										870	1127
2014	410										669	900
2015	332					511			1960		1654	2108
2016	878		399	496	878				282		1892	2494
2017	488	672	472	601	507	699	499	638	2351	3176	995	1285

Appendix 4.xii. Input data for 2SW, large and small salmon returns and spawners to USA used in the run-reconstruction.

	Returns of 2SW	Returns of large salmon	Returns of small salmon	Spawners of 2SW	Spawners of large salmon	Spawners of small salmon
	USA	USA	USA	USA	USA	USA
Year of return						
to rivers	Point estimate	Point estimate	Point estimate	Point estimate	Point estimate	Point estimate
Winbugs labels	USAR2[]	USALg[]	USASm[]	USAS2[]	USASLg[	USASSm[
1970	0	0	0	0	C	
1971	653	653	32	490	490	29
1972	1383	1383	18	1038	1038	1
1973	1427	1427	23	1100	1100	1;
1974	1394	1394	55	1147	1147	4
1975	2331	2331	84			6
1976	1317	1317	186			
1977	1998	1998	75	643	643	
1978	4208					
1979	1942					
1980	5796					
1981	5601	5601				
1982	6056					
1983	2155				1769	
1984	3222				2547	
1985	5529					
1986	6176				5570	
1987	3081	3081			2781	
1988	3286					
1989	3197					
1990	5051	5051				
1991	2647	2647				
1992	2459				2292	
1993	2231	2231				
1994	1346					
1995	1748					
1996	2407				2407	
1997	1611	1611			1611	
1998	1526					
1999	1168					
2000	533				1587	
2001	788				1491	
2002	504	526			511	
2003	1192					
2004	1283			1283	1283	319
2005	984	994	319	1088	1088	
2006	1023	1030	450	1419	1419	450
2007	954	958	297	1189	1189	297
2008	1764	1799	814	2809	2231	814
2009	2069	2095	241	2292	2318	24
2010	1078	1098	525	1482	1502	525
2011	3045	3087	1080	3872	3914	1080
2012	879				2054	
2013	525					
2014	334					
2015	761	771				
2016	389					
2017	663					

## Appendix 5: Model Walkthroughs

Summaries of the data preparation, model running and output processing were presented at a one-day workshop prior to the 2014 WGNAS meeting and provided step by step walkthroughs of the assessment processes. Where appropriate these have been updated in 2015.

# NEAC pre-fishery abundance and national conservation limit model in R

[NB: Instructions apply to model version: "NEAC\_PFA\_CL\_RR\_model\_2015-varM\_v12" as used in 2015]

### 1) Introduction

This program performs the run-reconstruction estimation of pre-fishery abundance (PFA) of maturing and non-maturing 1SW salmon for each country (and region) in the NASCO-NEAC area. PFA is estimated for January 1st in the first sea winter. The program also establishes the pseudo stock–recruitment (S–R) relationship between lagged egg deposition and Total 1SW PFA, and applies a hockey-stick S–R analysis to estimate the National/Regional Conservation Limit where river-specific CLs are not available. The original model is described by Potter *et al.* (2004), minor changes to the estimation approach used for different countries and regions have been reported in the annual reports of WGNAS.

The model also estimates the proportion (mean and SE) of the Faroes catch originating in different countries/regions based on the genetic analysis of scales collected in the fishery between 1993 and 1995 and the estimated PFA for each country/region since 2001, when no fishery has been operating at Faroes. This requires the model to be run once to provide the catch proportions and then run a second time to provide a full PFA assessment. The catch proportions are also used in the Catch Options model.

- 1) To get started
  - 1.1) Load RStudio or R;
  - 1.2) Set up a folder from which you will run the program;
  - 1.3) Use folder and file names without spaces;
  - 1.4) Put the program, the input files (annual and multiannual) and the summary data file (see 6f) in this folder.
- 2) Input Data
  - 2.1) Annual data (filenames: Annual-data-XX-YY.txt)
    - 2.1.1) There is a file for each country (XX) and region (YY) which contains the 40+ year time-series of data on catches, exploitation rates and non-reporting rates (plus additional data for some countries).
    - 2.1.2) To read the .txt files, it is easiest to open them from within Excel. i.e.
      - Open Excel;
      - select the correct folder;
      - click on 'Open'
      - You will probably need to change the setting in the lower right corner of the open box from 'Excel files' to 'All files';

• Double-click on the file you want to open and it should open the 'Text Import Wizard';

- select 'finish' (If this doesn't work reopen the file, but select 'Delimited' at step 1, 'Tab' at step 2 and 'General' at step 3.)
- 2.1.3 ) Do not add any formatting to the file. If loading a new version of a file that has been saved in Excel (e.g. after addition of a new year's data), re-save the file by clicking 'Save As' and selecting 'Text (Tab delimited)' from the 'Save as type' list. This will remove the formatting and add the .txt extension.
- 2.1.4) Do not change the file name.
- 2.1.5) Close and save the file before running the programme. You will be prompted to confirm that you want to lose the formatting; click 'yes'.
- 2.2 ) Multiannual-data (file-name: 'Multiannual-data.txt' or similar)
  - 2.2.1) This file contains most of the other parameters used in the model including: smolt age composition, fecundity and sex ratios by region, M, etc.
  - 2.2.2) The second value listed is the 'lastdatayear' which needs to be updated to the latest year for which data are provided in the Annual-data-XX files.
  - 2.2.3 ) The file is not formatted in columns so can be read easily in Notebook, which should be selected automatically if you click on the file to open it. (NB: If you open the file in Excel, don't save it because it will probably add "" marks to each line.
  - 2.2.4) All blank lines and lines starting with '#' are ignored in this file. Apart from these:
  - The first line must start with 'list(
  - The last line must be ')'
  - All other lines must be 'variable name' <- number, followed by a comer (except for the last data line which has no comer).
  - 2.2.5) If the module estimating the composition of the Faroes catch is run (see below) the new values must be inserted at the end of the multiannual data file in place of the current ones.
  - 2.2.6) Save the file before running the model.

# 3) Model structure

- 3.1) Introductory section: contains working directory, source files and various parameters controlling the way the program runs (some of these will need to be changed for your laptop (see below).
- 3.2) Functions: functions are sections of code that the program calls up to repeat the same job. They have to be run before they are first called by the program; this is achieved by placing them at the beginning of the code. The main functions run the hockey-stick analysis for the NCL model and output certain figures and tables.
- 3.3 ) Faroes and Greenland sections: these sections calculate the harvest in the distant water fisheries.

3.4) NEAC country/regions sections: there is a section for each country (in alphabetic order) and region to calculate the main outputs of the R-R model.

- 3.5) Output summaries: this section creates NEAC summary figures and tables and the country/region data files for the Winbugs Forecast Model.
- 3.6) Faroes catch composition: The final section estimates the proportion (mean and SE) of the Faroes catch originating in different countries/regions. This requires the model to be run once to provide the catch proportions and then run a second time to provide a full PFA assessment.
- 4) Running the code from RStudio
  - 4.1) Open R Studio
  - 4.2) Select "File/Open File" and use the browser to select and open the code file; the code should open in the Top left panel. The code will have a name like "NEAC\_PFA\_CL\_RR\_model\_2015\_xxx"
  - 4.3 ) If you have been using the code recently, you can select "File/Recent Files" and select the file from the drop-down list (if it is there); you can open several code files simultaneously and they appear as tabs above the Top Left panel.
  - 4.4) To set up the code for your PC/laptop, R-click on the code and scroll down to:

line 40 -enter the full path name of the working directory (replace the text between the parentheses with the full pathname of the folder containing the code on your laptop (e.g. "D:/Modelling\_NEAC/PFA\_NCL\_R/2014").

line 45 -ensure that the text between the parentheses shows the correct filename for the multi-annual data file.

lines 77-86 -select which countries you wish to run the assessment for by setting "run-XX": 1 = run country XX; 0 = do not run. The summaries will only be run if all countries are set to 1.

line 82 -set "PrintFigs" equal to '1' to output the summary figures (or any other value not to output them); otherwise "0".

line 89 -set "WinbugsFiles" equal to '1' to output the data files for the Bayesian forecast model (or any other value not to output them); otherwise "0"...

line 92 -set "PrintCountryTables" equal to '1' to output summary output data for each region that is run (or any other value not to output them); otherwise "0"..

line 98 -set "RunFaroeseCatchSplitEstimation" equal to "TRUE" to run the estimation of the Faroes catch composition; otherwise "FALSE".

- 4.5) You do not need to save your changes before you run the code. [If you wish to save any changes, use "File/Save" or "File/Save As" as normal. It's a good idea to include the extension ".R". NB: You will be prompted to save the file before you close it.]
  - 4.5.1) To clear the 'console' area (lower left panel) press "Ctrl-L"

- 4.5.2) To run the program press "Ctrl-Alt-R"
- 4.5.3) You will see when part of the code run in console area. Errors will show in red. The run is complete when the final line shows ">"
- 5) Running the program from R
  - 5.1) Open R Studio
  - 5.2) Select "File/Open script" and use the browser to select and open the code file; the code should open in a separate panel. The code is currently called "NEAC\_PFA\_CL\_RR\_model\_2014"
  - 5.3 ) To set up the code for your PC/laptop, R-click on the code and scroll down to:
    - # SET WORKING DIRECTORY (wd): In line starting "wd <-" replace the text between the parentheses with the full pathname of the folder containing the code on your laptop (e.g. "D:/Modelling\_NEAC/PFA\_NCL\_R/2014").
    - #SET "run\_XX": in the lines starting "run\_XX <- " select which countries you wish to run the assessment for by setting "run-XX": 1 = run country XX; 0 = do not run. The summaries will only be run if all countries are set to 1.
    - # SET 'PrintFigs': set "PrintFigs" equal to '1' to output the summary figures (or any other value not to output them).
    - # SET 'WinbugsFiles: set "WinbugsFiles" equal to '1' to output the data files for the Bayesian forecast model (or any other value not to output them).
    - # SET 'PrintCountryTables': set "PrintCountryTables" equal to '1' to output summary output data for each region that is run (or any other value not to output them).
    - # SET 'RunFaroeseCatchSplitEstimation': set "RunFaroeseCatchSplitEstimation" equal to "TRUE" to run the estimation of the Faroes catch composition; otherwise "FALSE" [SEE BELOW]
  - 5.4) You do not need to save your changes before you run the code, but you may wish to save a version to be safe. To do this use "File/Save" or "File/Save As" as normal. It's a good idea to include the extension ".R". NB: You will be prompted to save the file before you close it.
  - 5.5) To run the program select "Edit/run all"
  - 5.6 ) You will see when the code runs in the 'R console' panel. Errors will show in red. The run is complete when the final line shows ">"
- 6) Running the Faroes stock composition
  - 6.1) The 'Multiannual-data' file contains the latest estimates of the composition of the Faroes catch by European country/regions based on the results of the genetics analysis reported in the 2015 WGNAS report and the 2001-14 PFA outputs. These estimates may be updated if new genetics data are provided or additional years of PFA estimates are to be included.

6.2) To run the estimation, ensure that all "Annual-data-XX" files have been updated.

- 6.3) SET 'RunFaroeseCatchSplitEstimation' to "TRUE" and run the model.
- 6.4) The new stock composition parameters will be output in the file "Faroes\_split\_estimate.txt"; these data should be copied into the end of the "Multiannual-data" file to replace the values already there. These data are also required for the Catch Options models.
- 6.5 ) Reset SET 'RunFaroeseCatchSplitEstimation' to "FALSE" and run the model again to produce full updates of the PFA estimates.

# 7) Output files

The program produces the following outputs (if requested):

7.1) National plots: (filenames "Figure-XX")

PDF files showing the national plots currently used in the WG report. This includes: maturing and non-maturing 1SW PFA; returns and spawners for 1SW and MSW; homewater exploitation rates, and total catches (incl. non-reported) for each country (XX). It also shows the pseudo stock–recruitment hockey-stock plots for each region; these show the estimated CL, where this is used in the assessment.

7.2) Regional data: (filenames "Region\_data\_XX\_YY")

Excel files showing PFA, returns, catch, exploitation rates and spawners for 1SW and MSW fish and total eggs and lagged egg estimates for each country (XX) and region (YY).

7.3 ) Input files for Forecast analysis: (filenames: "Winbugs Data XX YY")

Excel files for each country/region containing mean and s.d. estimates for the simulations for lagged eggs, 1SW returns and MSW returns.

- 7.4) Summary tables by country:
- Median spawner numbers
- Conservation limits and SERs
- Maturing 1SW PFA
- 1SW returns
- 1SW spawners
- Non-maturing 1SW PFA
- MSW returns
- MSW spawners
- 7.5) Summary plot for N-NEAC and S-NEAC
- 7.6) A formatted Excel workbook is set up to link to the output files and format the tables ready for use in the WGNAS report.
- 8) Common problems
  - 8.1 ) The code will crash if an output file (Figure or Table) is left open. The error message (in red) may say:
  - Error .....: cannot open file 'Fig-XX'

• Error in .....: : cannot open file 'Region\_data\_XX.csv': Permission denied

- 8.2 ) It doesn't matter if an input file is open, but the program may not read the latest version if it has not been saved.
- 8.3) More problems to be added .... when they are found!

