## Stock Annex for Sea bass in IVb,c and VIIa, d-h

Stock-specific documentation of standard assessment procedures used by ICES.

| Stock | European sea bass (Dicentrarchus labrax) in Subarea <br> IVb,c and VIIa, d-h (BSS-47) |
| :--- | :--- |
| Working Group | WGCSE |
| Date | May 2014 |

Revised by Mike Armstrong and Mickael Drogou, May 2015

Revisions
v1. WGNEW (bass in all areas)
v2. IBPNEW 2012 (retaining only information for BSS-47)
v3. WGCSE 2013 with revisions
v4. WGCSE 2014 with revisions from IBP-bass
v5. WGCSE 2015

## A. General

## A.1. Stock definition

Bass Dicentrarchus labrax is a widely distributed species in Northeast Atlantic shelf waters with a range from southern Norway, through the North Sea, the Irish Sea, the Bay of Biscay, the Mediterranean and the Black Sea to Northwest Africa. The species is at the northern limits of its range around the British Isles and southern Scandinavia.
Stock structure of sea bass in the Northeast Atlantic has been reviewed by WGNEW 2012 and IBP-NEW 2012 based on evidence from genetics studies, tagging studies, distribution of commercial catches and similarities in stock trends between areas, drawing also on extensive information contained in previous WGNEW and ICES SGBASS reports.
IBP-NEW and WGCSE considers that stock structure remains uncertain, and recommends further studies on sea bass stock identity, using conventional and electronic tagging, genetics and other individual and population markers (e.g. otolith microchemistry andshape), together with data on spawning distribution, larval transport and VMS data for vessels tracking migrating bass shoals, to confirm and quantify the exchange rate of sea bass between sea areas that could form management units for this stock. Such information is critical to support development of models to describe the spatial dynamic of the species under environmental drivers (e.g. temperature and food). Such a modelling work is being carried out in France in the framework of a PhD study ( R . Lopez), and in the UK.

The pragmatic view of IBP-NEW 2012 was to structure the baseline stock assessments into four units:

- Assessment area 1. Sea bass in ICES Areas IVbc, VIId, VIIe,h and VIa,f\&g (lack of clear genetic evidence; concentration of Area IV bass fisheries in the southern North Sea; seasonal movements of bass across ICES divisions). Relatively data-rich area with data on fishery landings and length/age composition; discards estimates and lengths; growth and maturity parameters; juvenile surveys, fishery lpue trends. [Bass-47].
- Assessment area 2. Sea bass in Biscay (ICES Subarea VIIIa,b). Available data are fishery landings, with length compositions from 2000; discards from 2009; some fishery lpue.
- Assessment area 3. Sea bass in VIIIc and IXa (landings, effort).
- Assessment area 4. Sea bass in Irish coastal waters (VIa, VIIb, VIIj). Available data: Recreational fishery catch rates; no commercial fishery operating. [Bass-wosi].


Figure A1.1. Current stock definitions for sea bass.
Fishery landings of sea bass are extremely small in Irish coastal waters of VIIa and VIIg and the stock assessment for assessment area 1 will not reflect the sea bass populations around the Irish coast, which may be more strongly affiliated to the population in area 4 off southern, western and Northern Ireland.
At IBP-New (ICES, 2012a), it was agreed that sea bass in the North Sea (IVb\&c) and in the Irish Sea, Channel and Celtic Sea (VIIa,d,e,f,g\&h) would be treated as a functional stock unit as there is no clear basis from fishery data, tagging and genetics studies to subdivide the populations in the Irish Sea, Celtic Sea, Channel and North Sea into independent stock units. Supporting information can be found in the IBP-New report. The other stock units defined for sea bass are: west of Scotland and Ireland (VIa and VIb,j; Bay Of Biscay (VIIIa,b); and the more southerly population in VIIIc and IXa.

In the absence of new information the pragmatic view of WGCSE 2015 is to continue to assume the presence of discrete sea bass stocks off southern Ireland and in the Bay of Biscay (VIIIab) and Iberian waters (VIIIc, IXa).

## A.2. Fishery

## General description

The commercial sea bass fisheries in Areas IV and VII have two distinct components: an offshore fishery on pre-spawning and spawning bass during November to April, predominantly by pelagic trawlers from France and the UK, and small-scale fisheries catching mature fish returning to coastal areas following spawning and in some cases immature sea bass. The inshore fisheries include many small ( 10 m and under) vessels
using a variety of fishing methods (e.g. trawl, handline, longline, nets, rod and line). The fishery may be either targeting sea bass or taking sea bass as a bycatch with other species. Historical landings data for the small-scale fisheries have often been poorly recorded. Although sea bass can occur as target or bycatch of many vessels, the bulk of the catch can be taken by relatively few vessels. For example in the UK in 2010, sea bass landings were reported by 1480 vessels (including 1207 of 10 m and under), $10 \%$ of which were responsible for over $70 \%$ of the total landings of 719 t (Walmsley and Armstrong, 2012). For France, in 2009 sea bass landings were reported by 2226 vessels including 976 of 10 m and under. Three main métiers were responsible for over $83 \%$ of the total landings. Pelagic trawlers ( $31.5 \%$ of total landings, for 58 vessels and 276 seamen) and "liners+handliners" ( $21.7 \%$ of total landings for 416 vessels and 634 seamen) are very economically dependent of this species (Drogou et al., 2011). French bottom trawlers often do not target sea bass, but this gear does represent $30.1 \%$ of the total landings (for 832 vessels and 2769 seamen). (Drogou et al., 2011).

In 2014 a large decrease in French midwater landings has been observed due to the very bad weather conditions, and the vessels switched effort to hake. Seabass landings have also moved from 1591 tons in 2013 to 242 tons in 2014, and this métier accounted in 2014 for only $9 \%$ of the total international landings. A decrease was also observed in French bottom trawlers for the same reason to a lesser extent.

According to the CHARM 3 Atlas of the Channel Fisheries, sea bass production in value represented $€ 31937$ in 2008. It's the third most valuable species caught in the Channel (source: Agrimer) in 2008 behind sole and monkfish (tuna is not included in statistics). The market value sea bass depends greatly on how it's caught, giving added value to certain métiers: in 2013 according to the database SACROIS (Ifremer-DPMA), mean price of sea bass sold in France by liners was $€ 17$ per Kg compared with $€ 7$ per Kg for pelagic trawl, reflecting differences in volume and fish condition.
Sea bass are a popular target for recreational fishing in Europe, particularly for angling in the UK, Ireland and France, and increasingly in parts of southern Norway, the Netherlands and Belgium. Relatively little historical data are available on recreational fisheries although several European countries are now carrying out surveys to meet the requirements of the EU Data Collection Framework and for other purposes (ICES 2009; 2010; 2011, 2012c; Herfault et al., 2010; Rocklin et al., 2014; Van der Hammen and De Graaf, 2012,205; Armstrong et al., 2013.).

More detailed descriptions of national fisheries can be found in ICES WGCSE 2015 (ICES 2015) and in Armstrong and Drogou (2014).

Fishery management regulations
Sea bass are not subject to EU TACs and quotas. Commercial vessels catching bass within cod recovery zones are subject to days-at-sea limits according to gear, mesh and species composition. Under EU regulation, the MLS of bass in the Northeast Atlantic is 36 cm total length, and there is effectively a banned range for enmeshing nets of 7089 mm stretched mesh in Regions 1 and 2 of Community waters ${ }^{1}$. Emergency measures

[^0]were introduced by the EU banning pelagic / midwater trawling for bass from January to end of April 2015. A bag limit of 3 fish per day was introduced in 2015 for recreational fishing (EU Regulation 2015/523 of 25 March 2015).

A variety of national restrictions on commercial bass fishing are also in place. These include:

- a landings limit of 5 t /boat/week for all French and UK trawlers landing bass;
- closure of 37 bass nursery areas in England and Wales to specified fishing methods;
- UK regional byelaws in Cornwall and South Wales stipulating a 37.5 cm MLS;
- a minimum gillnet mesh size of 100 mm in South Wales;
- a variety of control measures in Ireland that effectively ban commercial fishing for bass in Irish waters; plus MLS of 40 cm ;
- a licensing system from 2012 in France for commercial gears targeting sea bass.
- voluntary closed season from February to mid-March for longline and handline bass fisheries in Brittany.

Depending on country, measures affecting recreational fisheries include minimum landing sizes, restrictions on sale of catch, bag limits (Ireland), and gear restrictions (France; Netherlands).

## Management applicable to 2015

In IVbc and VIIa,d-h (North Sea, Channel, Celtic Sea and Irish Sea) the European Council has adopted emergency measures to help sea bass recover (Recent scientific analyses have reinforced previous concerns about the state of the stock and advised urgently to reduce fishing by $80 \%$ ). Effective emergency measures in January 2015 placed a ban on targeting the fish stock by pair-trawling while it is reproducing, during the spawning season, which runs until the end of April 2015. Bag limit of 3 fish per day was introduced for recreational fishing in 2015. The European Commission is consulting with Member States and stakeholders during 2013-2015 to establish a broader, balanced package of measures to control fishing mortality and rebuild the stock.

## A.3. Ecosystem aspects

Temperature appears to be a major driver for bass production and distribution (Paw1992). Reynolds et al. (2003) observed a positive relationship between annual seawater temperature during the development phases of eggs and larvae of sea bass and the timing and (possibly) abundance of post-larval recruitment to nursery areas. In addition, early growth is related to summer temperature and survival of 0-groups through the first winter is affected by body size (and fat reserves) and water temperature (Lancaster, 1991; Pawson, 1992). Prolonged periods of temperatures below 5-6 ${ }^{\circ} \mathrm{C}$ may lead to high levels of mortality in 0-groups in estuaries during cold winters. As a result, any SSB-recruit relationships may be obscured by temperature effects (Pawson et al., 2007a). WGCSE 2014 presented time series of UK inshore coastal temperatures

Region 2: All waters situated north of latitude $48^{\circ} \mathrm{N}$, but excluding the waters in Region 1 and ICES Divisions IIIb, IIIc and IIId.
during January to March and examined correlations with bass recruitment (Figure A.3.1).
a. Mean Jan-March sea surface temperature at UK coastal sites


Figure A.3.1. (a) Mean January-March coastal sea temperatures at five sites along the south coast of England, compared with the time-series of sea bass recruitment estimates from the update WGCSE assessment, including pre-1985 values from estimated recruit deviations back to 1975. (b)-(d): correlation between sea bass recruitment and (b) mean temperature for the five sites in the year of spawning; (c) mean temperature in the year of spawning and the following spring; and (d) the same as (c) but restricting the series to 1988 onwards (From ICES WGCSE 2014).

## B. Data

## B.1. Commercial catch

## B1.1 Landings data

## Data available

Landings series for use in the assessment are available from three sources:
i) Official statistics recorded in the ICES FishStat database since around the mid-1970s.
ii French landings for 1999-2013 from a separate analysis by Ifremer of logbook, auction and VMS data, with earlier official data adjusted as d scribed below.
iii ) Survey estimates of landings from the UK fleet of 10 m and under vessels (which are not obliged to provide EU logbooks), carried out by Cefas.

Total international landings of sea bass in European waters from sources (i) and (ii) combined increased from around 2000 t in the late 1970s to over 8000 t by 2006, the bulk coming from Areas IVb,c, VIIe and VIII. An important driver of the increase in landings since the 1990s was the increased landings in Divisions IVb,c, VIId and VIIe,h, coinciding with the large 1989 year class and a northward expansion of the sea bass population in the North Sea during a period of increasing sea temperatures.

Quality of official landings data
The official landings data for sea bass available to WGCSE 2015 are subject to several uncertainties that can affect the accuracy of assessments:

- Incomplete reporting of landings in the 1970s and early 1980s when the fisheries were developing (the assessment uses only data from 1985 onwards);
- Reporting of official French data by port rather than fishing ground before

2000. (The best landings estimates are from auctions for this period. During WGCSE, no fishing grounds could be identified for these landings).

From 2000 onwards, Ifremer has provided revised French landings from a separate analysis of logbook and auction data which allocates landings correctly by fishing round. To generate a consistent series of French landings from 1985 onwards for the Area IV\&VII assessment, IBP-New 2012 adjusted pre-2000 official landings from the ICES database by the average of the Ifremer correction factors by area from 2000-2010:

- IVbc+VIId: 1.04; VIIeh: 1.6; VIIafg: 0.62.

The accuracy of UK landings statistics is expected to have improved since the introduction of the Registration of Buyers and Sellers scheme in 2005/2006, particularly for recording of marketed landings of small vessels that do not have to supply EU logbooks. However, the official reported landings of sea bass in the UK are known to underestimate the true total landings, particularly for small-scale inshore fisheries where there has been no requirement to submit EC logbooks. Prior to the introduction of Buyers and Sellers requiring sales documentation, local fishery inspectors estimated landings of the under- 10 m fleet using whatever information they had available from
auctions, and frequently entered aggregated estimates for multiple vessels into the fishery landings database. Unfortunately the Buyers and Sellers regulations do not cover all landings. The EU Control Regulation allows landings of less than 30 kg to be disposed of for personal consumption without providing sales slips or other documents. This is ostensibly to reduce the administrative burden for a skipper disposing of small quantities for personal use. However, for small-scale fisheries where there are very large numbers of small vessels often catching small quantities, the cumulative catch of unrecorded small landings can be relatively high. This is likely to be an issue over the full time-series.

The UK(England) has previously carried out independent surveys to estimate historical landings data for sea bass, particularly for smaller vessels not supplying EU logbooks. A voluntary logbook scheme was carried out in conjunction with a biennial census of vessels catching sea bass. The census covers different segments of coast in different years (Pickett, 1990). The scheme was stratified by area, gear and vessel characteristics. Selected vessels from the strata kept logbooks for periods ranging from one to 25 years, and comprised what could be described as a "reference" fleet as opposed to a randomised selection of vessels each year. The scheme was terminated in 2007 and 2008, and reinstated for a further two years (2009 and 2010) before being terminated again. The scheme has now been suspended permanently. The landings tables in some earlier ACOM advice included "unallocated" landings which were the difference between the voluntary logbook estimates and the official UK statistics in each ICES area.
A review of the Cefas logbook scheme in 2012 (Armstrong and Walmsley, 2012a) showed that the previous estimates included recreational charter boats. After removal of these landings, the Cefas logbook estimates for nets and lines still showed substantial differences with official estimates (Figure B1.1). For fixed/driftnets, the landings including the Cefas logbook estimates for under 10 m vessels results in a landings series that is on average around three times higher than the official statistics. For lines, the ratio fluctuates around 3.0 for a large part of the series but was larger from 20002005. Insufficient logbooks were available for trawls to allow estimation of fleet-wide landings.
The Armstrong and Walmsley (2012a) review concluded that the survey is sensibly spread over a range of vessel types and gears, but is over-stratified and has insufficient (and declining) coyerage of the many survey strata while using adhoc, judgment-based vessel selection schemes rather than randomized selection. Despite the potential biases, the survey results for commercial vessels confirm that the historical official reported landings of sea bass are likely to be underestimates. Neither data source for UK 10m and under vessels is considered accurate over the historical time series but ICES WGNEW (ICES, 2008) and IBPBass (ICES, 2014) found that the stock trends from a statistical assessment model were relatively insensitive to the choice of these two catch histories. The main effect of including the additional landings is to scale biomass estimates up without altering the trend. Total fishing mortality estimates are not affected, as the age profiles of the catches remains more-or-less the same, but the proportion of total F attributable to recreational fisheries in the final stock assessment is reduced slightly by the additional commercial landings. Given the small contribution of UK reported landings of under- 10 m fleets to total international landings of bass, the uncertainties concerning the level of bias in the Cefas logbook estimates, and the lack of such estimates after 2011, the official statistics have been retained in the current assessment. Of more concern would be any change in reporting accuracy over time, as this would lead to biased assessment trends using reported landings figures.


Figure B1.1. Top: estimates of sea bass landings for under-10 m UK netters and liners based on the Cefas logbook and port survey scheme. Bottom: ratio of landings of these gears including the Cefas logbook estimates for $<10 \mathrm{~m}$ commercial vessels, and the total official reported landings of all UK vessels using these gears.

## Dutch landings

Landings and effort data from the commercial fleet are available from the EU logbooks; market category composition of landings is available from the auction data (sale slips); and size and age data are available through market sampling. The fisheries were described in detailin van der Hammen et al. (2013).

## EU logbook data

Official EU logbook data of the entire Dutch fleet are maintained by the NVWA (formerly known as the General Inspection Service, AID). IMARES has access to these logbooks and stores the data in a database (VISSTAT). EU logbook data contain information on:
landings (kg): by vessel, trip, ICES statistical rectangle and species;

- effort (days absent from port): by vessel, trip and ICES statistical rectangle;
- vessel information: length, engine power and gear used.

Logbook data are available from the entire Dutch fishing fleet and from foreign vessels landing their catches in the Netherlands.

## Auction data: landings by market category

Auction data cover both the total Dutch fishing fleet and foreign vessels landing their catches on Dutch auctions. These data are also stored in VISSTAT and contain information on:

- landings by market category (kg): by vessel, trip (landing date) and species.

Further information on availability and quality of landings data by country is provided by IBPBass (ICES, 2014).

## B1.2 Discards estimates

## UK data

## Survey design and analysis

Estimation of bass discards is problematic because the observer scheme covers all vessel trips in a stratum without reference to target species, and overall it samples less than $1 \%$ of all fishing trips. As bass are absent or at very low numbers in a large fraction of fishing trips throughout the year, particularly in winter, the amount of sample data on bass is very low and the estimates are likely to have poor precision and variable biases related to inclusion of under 10 m vessels. Vessels under 10 m , that are responsible for a large fraction of bass catches, were excluded until recent years on health and safety and other logistical grounds, and although under 10 m vessels are now included since 2007, the fleet of vessels under 7 m remains excluded.
Raising of UK discards data to the fleet level is currently performed using a ratio estimator using the ratio of reported landings in a stratum to the computed retained weight of sea bass in sampled trips. This assumes that the retention curve for sampled vessels is representative of all vessels in the stratum.

## Data coverage and quality

UK discards data are available for métiers associated with trawls and fixed/driftnets only. Discards from commercial line boats are expected to be relatively low and have high survival, so this fleet sector is excluded from the scheme for sea bass. As sampling is targeted at all species, annual coverage of the bass fisheries is relatively limited.

Numerically, the largest numbers of bass discarded from UK fisheries are from the bottom trawl fleet, with much smaller numbers from nets and even less from beam trawls. Only eleven midwater pair trawl trips were observed up to 2013, and discarding of bass was negligible as the fishery targets mainly adult bass. No bass discards were observed in the eight longline trips observed. The raised length frequencies of discarded and retained bass, aggregated over all years, are available along with the retention ogives. It is clear that discarding is driven primarily by the minimum landing size of 36 cm (see IBPBass 2014 report).
Sampling levels for UK and French discard sampling, and discards estimates, are given in the WGCSE report and updated annually.

French data

## Survey design and analysis

The French sampling schemes also utilise vessel-list sampling frames and random selection of vessels within strata defined by area and fleet sector. From the activity calendars of French vessels for year n-1, vessels are grouped by the métiers practiced. Thus, a vessel may belong to multiple groups if practicing several métiers in the period. If the métier has to be sampled in priority No. 1, the vessel to be boarded is chosen randomly within this group of vessels. The observer then chooses to go on board for a trip. During the trip, the fishing operations corresponding to métier No. 1 are sampled. Optionally, if the vessel practices several métier during the trip, fishing operation of
the métier No 2 will also be sampled if the métier No. 2 is included in the annual sampling plan. If the métier is not part of the plan, it is requested to sample at least one fishing operation of this métier in the trip. (complete document on sampling protocol in French :http://sih.ifremer.fr/content/download/5587/40495/file/Manuel OBSMER V2 2 2012.pdf)

## Data coverage and quality

Discards data are only available for French fleets from 2009 onwards. For 2012, results are described in the annual French report on:" http://archimer.ifremer.fr/doc/00167/27787/25978.pdf" for 2012. Discards estimates for France are from vessel selections that for some areas and gears include relatively limited numbers of observed trips where sea bass is caught and discarded. Precision is therefore very low at current sampling rates. In France the low sampling rate observed can be explain by the low discarding rate. Length frequencies were not available at IBPbass for 2013.

## Spain

No bass discards were observed for any métier in the 2003-2013 period. Number of sampled hauls per métier and area were presented to IBP-NEW 2012 (see assessment report).

Discards data from other European countries
Discards data for Dutch beam trawlers were presented to ICES IBP-NEW 2012, as annual mean numbers discarded per hour in 2004-2010. No commercial fisheries for sea bass exist in Ireland.

## B1.3 Recreational catches

Recreational marine fishery surveys covering different parts of the sea bass stock in the North Sea, Channel, Celtic Sea and Irish Sea have been developed in France, Netherlands, England and Belgium (ICES, 2012c). Survey design and methods of recreational catch estimation are described in IBPBass 2014.

Data from France
A survey of recreational fishers, focusing mainly on bass, was conducted between 2009 and 2011. Estimates of sea bass recreational catches were obtained from a panel of 121 recreational fishermen recruited during a random digit dialling screening survey of 15000 households in the targeted districts. The estimated recreational catch of bass in the Bay of Biscay and in the Channel was 3170 t of which 2350 t was kept and 830 t released. The estimates for Subarea IV and VII were 940 t kept and 332 t released.

The precision of the combined Biscay and Channel estimate was relatively low (CV=26\%; note that the figure of 51\% given in IBPNEW 2012 (ICES, 2012a) was incorrect). This gives an average and $95 \%$ confidence intervals of 3170 t [1554 t; 4786 t ] for the whole Subareas IV, VII and VIII. Increasing the panel from 121 to 210 fishermen would be expected to improve precision to $20 \%$ and increasing this panel to 500 would improve precision to $13 \%$.

The main gears used, in order of total catch, were fishing rod with artificial lure, fishing rod with bait, handline, longline, net and spear fishing. Approximately $80 \%$ of the recreational catch was taken by sea angling (rod and line or handline).

A new survey was conducted from July 2011 to December 2012, based on a similar methodology to the previous study (not only on sea bass this time, but also on other marine species including crustaceans and cephalopods). A random digit dialling screening survey of 16130 households led to the recruitment of a panel of 183 fishermen to keep logbooks. In parallel, 151 fishermen were recruited on site by the Promopeche association, and 30 more via the sea bass fishermen panel set up in 2009. This resulted in 364 panel members keeping logbooks describing their catches (species, weight, size, etc.) The focus of the survey on sea bass shows that in Atlantic (Bay of Biscay and Channel), the estimated recreational catch of bass in 2012 was 3922 t of which 3146 t was kept and 776 t released. At this time results have to be considered as provisional, (results split between Bay and Biscay and Channel are not available yet with relative standard error).

## UK (E\&W)

A new survey programme Sea Angling 2012, based on a statistically sound suryey design started in 2012 to estimate fishing effort, catches (kept and released) and fish sizes for shore based and boat angling in England. The survey does not cover other forms of recreational fishing. Results are available in Armstreng et al., 2013.
The surveys adopted, where possible, statistically-sound, probability-based survey designs, building on knowledge gained through participation in the ICES Working Group on Recreational Fishery Surveys (WGRFS). Two survey approaches were adopted: firstly a stratified random survey of charter boats from a list frame covering ports in England, and secondly an on-site stratified random survey of shore anglers and private boat anglers to estimate mean catch per day, combined with annual effort estimates derived from questions added to a monthly Office of National Statistics household survey covering Great Britain.
A list of almost 400 charter boats was compiled for the charter boat survey, and 166 skippers agreed to participate. Each month over a twelve month period in 2012 and 2013, 34 randomly-selected skippers completed a diary documenting their activities, catches and sizes of fish. A diary was completed whether or not any fishing took place. Data from 5300 anglers were collected. Total annual catches were estimated by raising the monthly catches per vessel from the diaries to all vessel-month combinations in the frame, and raising this to all vessels including refusals. The estimated total annual catch of sea bass for the entire coast of England was 44 t (RSE $31 \%$ ) of which 31 t was kept. The release rate by number was $37 \%$. The charter boat survey has potential bias due to the large non-response rate, if non-respondents have different catch rates to respondents.
The Office of National Statistics (ONS) household survey covered 12000 households during 2012, and from this it was estimated that $2.2 \%$ of adults over 16 years old went sea angling at least once in the previous year. The surveys estimated there are 884000 sea anglers in England. Estimation of fishing effort by shore and private boat anglers proved very difficult due to the overall low number of households with sea anglers in the survey. A range of methods was explored to estimate annual and seasonal effort using the ONS data alone, and combining it with observations from on-site and on-line surveys. It has not been possible yet to agree on a best estimate of effort, and for that reason the estimates of total catch (cpue $\times$ effort) for shore and private boat angling are given as a range of plausible values.

The survey of anglers fishing from the shore and private boats to estimate cpue was carried out throughout 2012 using on-site interviews. A stratified random design was
adopted to select shore sites and boat landing sites on a weekly basis from site lists stratified into low-activity and high-activity sites. The shore survey used roving-creel methods (collecting data from partial angling trips), and the private boat survey a roving access-point survey (data from completed trips). Visits were made to 1475 shore sites and 425 private boat sites, and 2440 anglers were interviewed. The mean daily catch rate of kept and released fish of each species was estimated based on the survey design, and sizes of caught fish were recorded. Cpue for shore angling was estimated using catches for the observed trip duration and estimates of expected total trip duration for that day. A length-of-stay bias correction was applied based on expected total trip duration. The catch-per-day estimates were combined with estimates of total annual fishing effort (days fished) obtained from the ONS survey to estimate total annual catches. Release rates, by number, were $82 \%$ for shore angling and $57 \%$ for private boats. Non-response rates were very low ( $<10 \%$ ) in this survey. The range of point estimates for shore-caught bass was 98-143 t (total) and 38-56 t (kept), and for private and rented boats was 194-546 t (total) and $142-367 \mathrm{t}$ (kept). The relative standard errors for the individual shore and private boat estimates were relatively high at $40-50 \%$.
Combining the catch estimates for charter boats, private boats and shore angling, the point estimates of annual kept weights of sea bass ranged from $230 \mathrm{t}-440 \mathrm{t}$ (Table 2.4.1c), compared with total UK commercial landings of almost 900t in 2012. The combined estimates of bass catches had precision (relative standard error) estimates of $26 \%-38 \%$ for the different effort estimation methods. The relatively large standard errors combined with the range of plausible methods of estimating effort for shore and private boats.

## Netherlands

Sea bass are taken by recreational sea anglers in the Netherlands. A recent survey investigated the amount of sea bass caught by recreational fishers (van der Hammen and de Graaf, 2012, 2015; ICES, 2012c) from March 2010 to February 2011 and from March 2013 to February 2013. Estimates of recreational sea bass were obtained from a panel of 1043 recreational fishermen (in the first survey) recruited during a telephone survey of 109293 people. Revised estimates were provided to WGCSE 2013 (ICES, 2013a). The catch weights are estimated with a limited amount of length-frequency data, and are therefore less reliable than the estimates in numbers (and may also be adjusted if more data are available). For the same reason, there are no 'returned' estimates by weight (yet).
The estimated total recreational catch of sea bass was 366000 fish (RSE 30\%), of which 234000 were retained, equivalent to 138 t . These results are mainly applicable to SubIV.

## Belgium

A recreational fishing survey was conducted in 2013 in Belgium by the Belgian Fisheries Institute, using a questionnaire approach, in order to meet DCF requirements. The estimated retained catch of sea bass was 60 t .

## Hooking mortality rates

The US National Marine Fisheries Service has in the past used an average hooking mortality of 9\% for striped bass, estimated by Diodati and Richards (1996). Striped bass are very similar to European sea bass in terms of morphology, habitats and angling methods. A literature review of hooking mortality for a range of species compiled by
the Massachusetts Division of Marine Fisheries included a total of 40 different experiments by 16 different authors where striped bass hooking mortality was estimated over two or more days (Gary A. Nelson, Massachusetts Division of Marine Fisheries, pers. comm.) The mean hooking mortality rate was 0.19 (standard deviation 0.19). Direct experiments are needed on European seabass to estimate hooking mortality for conditions and angling methods typical of European fisheries.

## Total recreational catch

Total catches are given in the WGCSE report and updated when new estimates become available. They are also documented in the annual reports of the ICES Working Group on recreational fisheries (WGRFS).

## B.2. Biological sampling

## B2.1 Length and age compositions of landed and discarded fish in commercial fisheries

Length and age compositions of sea bass landings were available to WGCSE 2015 from sampling in the UK and France.

UK

## Sampling methods and analysis

The UK(E\&W) sampling programme for length compositions of sea bass covers sampling at sea and on shore. The sampling design for at-sea sampling is described above. The on-shore sampling programme uses an area list frame comprising port days, currently stratified by quarter, ICES Division and an index of "port size". "Large" ports are sampled more intensively than "small ports". Separate list frames of ports are established for pelagic trawlers, beam trawlers and demersal trawl, nets and lines. Sampling targets are set to achieve a specified number of port visits by stratum, taking account the need for fleet based as,well as stock based data specified by the EU Data Collection Framework, although other diagnostics are monitored such as numbers of fish measures and otoliths/scales collected by species. This scheme has only been in development and operation since around 2010 when Cefas took over the sampling from the Marine and Fisheries Agency. Prior to then, the sampling targets were mainly set as numbers of fish of each species to measure or age by quarter, district, and gear groupings, with ninimum numbers of sampling trips also specified to spread the sampling out.
Length compositions are first vessel-raised using ratios of landed live weight to predicted live weight of the length frequency calculated from a length-weight relationship:

$$
W(\mathrm{~kg})=0.00001296\left(\mathrm{~L}+0.5^{5}\right)^{2.969}
$$

Raised LFDs are then summed over vessels within a sampling stratum and raised to give total raised fleet LFDs per stratum, which are then combined. This procedure ensures sums-of-products ratios of 1.0, but will lead to some bias in numbers-at-length due to discrepancies between true fish weights and calculated fish weights from the length-weight relationship.

## Data coverage and quality

Age compositions for UK commercial fishery landings of sea bass are derived from biennial (January-June and July-December) age-length keys (ALK) constructed for
four areas: IVb,c; VIId, VIIe,h and VIIa,f,g. These are applied to fleet-raised landings length frequencies for each of four gear groups (bottom trawls; midwater trawls; fixed/driftnets and lines) in each area. Further details are given in the ICES IBP-NEW (ICES, 2012a) and WGCSE (ICES, 2013a) reports and in the stock annex along with tables giving numbers of trips sampled for length and age and numbers of fish measured and aged.
A recommendation of WGCSE 2013 was to expand the UK age frequencies to the full recorded age range and to re-evaluate the plus-group definition (previously at 12+). Sea bass have been recorded to almost 30 years of age, and it was thought that having more true ages in the Stock Synthesis input data could allow better estimates of early recruit deviations. The necessary extractions were done for IBPBass, and the data were examined in detail by Armstrong and Readdy (IBPBass Working Document_01). Bubble plots and catch curves showed that coherent information on year classes was present well beyond the last true age (11) previously adopted. The exploratory SS3 runs show that the different choices of plus-group (12+; 16+; 18+; 20+) have relatively little impact on the results, other than (as hoped) a slightly better estimation of early recruit deviations. Expanding the age compositions may help fit domed selection curves for fleets where this is appropriate, but risks an increasing humber of zero catch entries for older ages as recent weak year classes feed into future catches and become depleted. A plus-gp of $16+$ was recommended for further model development and agreed by IBPBass. Sampling of midwater trawls prior to 1996, and in 1997, was considered too poor to develop age compositions. All datasets show a very strong 1989 year class and very weak 1985-1987 year classes.

## France

## Sampling methods and analysi;

The French sampling programme for length compositions of sea bass covers sampling at sea and on shore. Since 2009, both sampling types are first based on métiers composition and their relative importance per fishing harbours and month. Both are also designed to sample the whole catch following a concurrent sampling of species, potentially leading to low sea bass sample size. In order to complement this effort, specific sampling for seabass at the market is added at times and harbours when higher landings are occurring, especially from métiers targeting sea bass. The sampling frame is based on the main harbours, gear types (or grouping of métiers) and month and is ayailable to all samplers on a dedicated website. Real time follow-up of the plan, refusal rates and their reasons, time taken to sample, all this information is also available from the website, together with sampling protocol (in French:
http://sih.ifremer.fr/content/download/5587/40495/file/Manuel OB-
SMER V2 2 2012.pdf). Before 2009, only market specific sampling was in place, and the sampling plan was designed and followed by the stock coordinator. The French sampling programme for age compositions of sea bass is based on age-length keys with fixed allocation. For the VIIeh area, quarterly French landings at auctions are sampled in order to collect five scales (from 2000 to 2008) or three scales (from 2009) by length class (cm). For the VIIIab area the information is available only from 2010. For other areas the information is not available. All length samples are populated in a central database (Harmonie) and regular extracts are available in the COST format. Raising the data to the population is done using COST tools and a special forum for discussing the outcomes of the analysis is held every year in March, in order to gather all stock coordinators and prepare the datasets for the assessment working groups.

## Data coverage and quality

Length compositions of French sea bass commercial landings are constructed from2014 for the area including IVb,c; VIId, VIIe,h and VIIa,f,g, for all métier. The input data for French fleets in Stock Synthesis are the fleet-aggregated length compositions in 2 cm classes (20-21.9, 22-23.9, etc.) for each year from 2000 onwards.

The statistical design of fishery sampling schemes has undergone change in recent years in the UK and France, following recommendations from ICES workshops on sampling survey design, with a move towards more representative sampling across trips within fleet segments. This can result in sampling more trips that have small catches of bass, and is one reason for the increase in numbers of sampled trips with bass since 2009 in France which does not imply an increase of the proportion in numbers of fish measured per trip.

## Other countries

Fishery landings length or age compositions from other countries catching bass were not available to IBPBass 2014. The Netherlands did collect age samples of sea bass every year from 2005 to 2008. From 2010 onwards, age samples are collected only once every three year. In the IMARES market sampling data on length, age, sex and weight are collected for several commercially important species. For sea bass this is done on an irregular basis and data are only available for some years. Market sampling is done since 2005. The age sampling frequency is now set triennially (2010, 2013, etc.) Every three years four samples of 15 fish ( 60 fish in total) are aged, and every year the lengths of 24 samples of 15 fish ( 360 fish in total) are taken. Otoliths and scales that are retrieved from the fish are sent to Cefas in the UK for age reading. Length samples are collected every year. All samples are collected in the auctions where most sea bass is landed, in the south of the Netherlands. The quality of the data is good enough to use them in assessments. However, bath the length and age data need processing before they can be inserted in an assessment.

## Effective sample sizes for length and age compositions

The effective sample size for annual estimates of length or age composition lie between the number of trips sampled and the number of fish measured or aged, due to cluster sampling effects. Effective sample sizes have not been computed yet for UK and French sampling data for sea bass. In the meantime, ESS for the combined UK trawls, nets and lines fleet are set at 100 for each year (20 for 1985-1990), and for the French combinedfleet LFDs, the ESS is set at 200 for each year.

Accuracy and validation of age estimates
Age-reading consistency
Consistency in age reading of sea bass between four operators in Cefas and Ifremer was examined during a limited exchange of otolith and scale images between laboratories in 2011, organised by the ICES Planning Group on Commercial Catches, Discards and Biological Sampling (Mahé et al., 2012). A total of 155 fish of $17-74 \mathrm{~cm}$ was sampled on board French research vessels during two international surveys. The precision of ageing was similar for scales and otoliths. The coefficient of variation of age readings for individual fish was around $12 \%$ implying a standard deviation of $+/-1$ year for a ten year-old fish, with relatively few fish having identical readings by all four operators. However it was noted by the operators that photographic images were more difficult to evaluate than original age material, which was likely to have a negative effect
on the consistency of ageing. These results provide no indication of the validity of ages, only the consistency between operators, and cannot indicate data quality in earlier years when different operators provided the age data. A more extensive age exchange is to be carried out in 2015.

## Age validation

WGCSE and IBPBass were not aware of specific studies to validate absolute ages of sea bass derived from otolith or scale readings. Strong and weak year classes can be followed clearly to over 20 years of age in UK sample data although it is not known to what extent the elevated numbers of sampled fish in immediately adjacent year classes is a true reflection of year-class strength or a consequence of age errors discussed in the previous section. Year-class tracking is less clear in the younger ages $3-5$ although this will be affected by gear selectivity and changes in fish behaviour.
Sea bass show relatively broad length-at-age distributions, and it has been noted in French data (Laurec et al., 2012 WD to IBP-NEW) that the length-at-age distributions can have unusual patterns including some multiple modes that could indicate age errors. This will result in some smoothing of age data across neighbouring year classes. In the UK data, unusual patterns in length-at-age distributions for some younger ages appear related more to effects of minimum landing size ondata from the fishery.

## Inclusion of age error parameters in Stock Synthesis model

CVs for ageing error by age class can be input to Stock Synthesis. Based on the ICES sea bass scale exchange in 2002, the CVs of $-12 \%$ can be specified as increasing values per age class to give a standard error of $\sim 1$ year per age class.

## B2.2 Growth parameters

Pickett and Pawson (1994) provide plots of growth curves for female and male bass based on samples collected in the 1980s in Areas IV and VII. The samples used by Pickett and Pawson (1994) for growth and maturity analysis were obtained from a range of fishery and other sources.

A re-analysis of UK historical age-length data including more recent samples was conducted in 2012, using data for the full UK sampling series from 1985 to 2010 (Armstrong and Walmsley, 2012b). The data are derived from sampling of UK fishery catches around England and Wales as well as from trawls surveys of young bass in the Solent and Thames estuary. More than 90000 sea bass have been aged since 1985. The inshore surveys are mainly young sea bass up to 3-5 years of age, whereas the fishery samples include fish up to 28 years of age.
Alkageing is done from scales, excluding scales considered to be re-grown. On surveys, scales are collected in a length-stratified manner from individual hauls with a view to building age-length keys. A similar approach has historically been adopted for catch sampling. This may lead to non-random sampling of individual age groups when the catch numbers are well in excess of numbers sampled from an individual catch. It will also lead to some overestimation of the standard deviation of lengths-at-age.

All ages for fitting growth curves are referred to a nominal January 1 birthdate, according to month of capture. Parameters of the von Bertalanffy growth curve were fitted in Excel Solver using non-linear minimisation of $\sum(\text { obs-exp })^{\wedge} 2$ for lengths-at-age of individual fish, by area and for all data combined.

Von Bertalanffy model parameters were as follows:

| AREA | IVBC | VIId | VIIe | VIIAFG | ALL AREAS |
| :--- | :---: | :---: | :---: | :---: | :---: |
| $\operatorname{Linf}(\mathrm{cm})$ | 82.98 | 87.22 | 92.27 | 81.87 | 84.55 |
| $K$ | 0.1104 | 0.09298 | 0.07697 | 0.09246 | 0.09699 |
| $\boldsymbol{t}_{0}$ (years) | -0.608 | -0.592 | -1.693 | -1.066 | -0.730 |

Standard deviation of length-at-age distributions increases linearly with age according to:

$$
\text { SD }(\text { age })=0.1166^{*} \text { age }+3.5609
$$

The sampled sea bass show some sexual dimorphism of growth from about seven years of age onwards. It is currently not possible to implement a sex-disaggregated Stock Synthesis assessment, therefore a combined-sex growth curve is adopted. Mean length-at-age has not shown any trend over time, and length-at-age is also very similar in strong and weak year classes (Armstrong and Walmsley, 2012b). Hence data have been combined over the full series to estimate growth parameters.

## B2.3 Maturity

## Spawning grounds and season

Ripe adult bass have been caught by pelagic trawling in the south of Division VIIIa and in the north of Division VIIIb in the Bay of Biscay during January-March (Morizur, unpublished data), and planktonic egg surveys (Thompson and Harrop, 1987; Jennings and Pawson, 1992) have shown that bass spawn offshore in the English Channel and eastern Celtic Sea from February to May. Spawning started in the Midwestern Channel when the temperature range associated with bass egg distributions was $8.5-11^{\circ} \mathrm{C}$, and appeared to spread east through the Channel as the surface water temperature exceeded $9^{\circ} \mathrm{C}$. Seasonal patterns of occurrence of advanced maturity stages in UK samples also indicate spanning mainly January to May in ICES Areas IV and VII (Armstrong and Walmsley, 2012c). Spawning and ripe bass are also found in the southern North Sea (information from commercial fisheries and angler reports in Netherlands supplied to IBP-NEW 2012 by F. Quirijns).

Previous estimates of maturity at length/age, and data available for re-analysis
BASS (ICES 2004) reported that around Britain and Ireland, male bass mature at a length of 31-35 cm, aged 4-7 years, and females at 40-45 cm, aged 5-8 years, (Kennedy and Fitzmaurice, 1972; Pawson and Pickett, 1996), and data from the southern part of the Bay of Biscay (Lam Hoai, 1970; Stequert, 1972) indicate that male matures at a length of 35 cm (age 4) and females at 42 cm (age 6). Data provided by Masski (1998) from samples taken from VIIe bottom trawlers ( 41 females) indicate that $40 \%$ and $82 \%$ of females were mature at-age 6 and 7 respectively, with a very small percentage mature at-age 5.

Collection of maturity data are difficult as few adult bass are caught in surveys and bass are typically landed whole and are extremely expensive to purchase. Samples collected by the UK (Cefas) during 1982-2003 and 2009 in ICES Areas IV and VII were reanalysed for ICES IBP-NEW 2012 (Armstrong and Walmsley, 2012c). Samples have
come from all around the coast of England and Wales, though few fish have been sampled in the Irish Sea (VIIa).

## Defining a maturity marker for sea bass

Sea bass are multiple batch spawners, as indicated by size distributions of oocytes (eggs) in ovaries (Mayer et al., 1990). This means that the ovary will start to mature oocytes through to vitellogenic stages during the months immediately prior to the spawning season. Historical maturity staging of sea bass by the UK has used the maturity key given in Pawson and Pickett (1996; Table B2-1). In their analyses, they treated stage 2 as mature, and stage 3 as immature. Their reasoning was that stage 3 ovaries (early maturing) were found in smaller bass than later stages (4+) indicating that many of these fish may not proceed to spawning. Sea bass migrate offshore to spawning grounds, and it is likely that early maturing fish could be over-represented, and advanced maturing fish under-represented in inshore catches sampled during the period of spawning migrations. An additional spent stage (VIII) has been occasionally recorded.

The identification of a suitable marker to identify maturity has to take into account the probability of finding a fish at any maturity stage in different months, the duration of a stage, and the availability/catchability of fish at that stage of maturity. When the majority of mature sea bass have entered the batch spawning cycle in spring, all stages represented in batch spawning (III to,VII) will be evident and should be distinct from immature fish. Hence, the best markersfor maturity are the maturity stages representing different stages in the batch spawning cycle, sampled at a time when spawning is taking place (or immediately before), provided fish in all stages are equally catchable. This is the conclusion of recent ICES workshops on maturity staging of gadoids and flatfish, which recommends sampling within a month or so of the beginning and end of the spawning season. Experience with other roundfish and flatfish stocks is that it can be very difficult to distinguish between virgin females and fish that have spawned previously, when sampled in the non-spawning period. The UK data were therefore re-analysed using samples from December to April, treating all fish of maturity stages 3 to 7 as mature.

## Re-estimation of maturity ogives from UK data

Maturity was modelled using a binomial error structure and logit link function, fitted in $R$ to individual observations. The logistic model describing proportion mature by 1 cm length class $L$ was formulated as:

$$
\operatorname{Pmat}(L)=1 /\left(1+\mathrm{e}^{-(a+b L)}\right)
$$

defined by the parameters slope $b$ and length intecept $a$. These parameters were estimated separately for females and males.This can also be expressed as:

$$
\operatorname{Pmat}(L)=1 /\left(1+\mathrm{e}^{-\mathrm{b}(L+\mathrm{c})}\right) \text { where } \mathrm{c}=\mathrm{a} / \mathrm{b}
$$

Stock Synthesis uses the second formulation, and the parameters required are the slope ( $b=0.3335$ : entered as a negative value) and the length inflection, which is the estimated length at $50 \%$ maturity $\left(\mathrm{L}^{50 \%}=40.65 \mathrm{~cm}\right)$.

The 2009 data come from a large sample of sea bass taken in spring from a few trips specifically to revisit bass maturity, but this sample dominates the time-series of sampling which is spread over very many more trips and months than in 2009 and therefore has better coverage. Maturity ogives were therefore fitted including and excluding 2009 data. The inclusion of 2009 data, which were for a relatively restricted length
range of fish around 40 cm , has the effect of improving the fit of the model near the top of the ascending limb of the maturity ogive for females (Figure B2-1). However the very high weighting for these lengths compared to the data for lengths $<35 \mathrm{~cm}$ results in the model fitting very poorly to the smaller length classes. Excluding the 2009 data allow the length classes $<35 \mathrm{~cm}$ to carry more weight, and the ogive appears to fit the data for $30-40 \mathrm{~cm}$ sea bass more closely, although the fit for lengths $>40 \mathrm{~cm}$ is poorer. Addition of the 2009 data effectively shifts the $\mathrm{L} 50 \%$ from around 41 cm to 35 cm . In contrast, inclusion or exclusion of the 2009 data has less effect on the model fit for males (Figure B2-1). On balance, it was considered undesirable for a few large hauls in a recent year to have excessive leverage in the model fit, and the model excluding 2009 was considered preferable as a long-term maturity ogive for use in assessments.

Table B2-1. Macroscopic characteristics of the maturity stages of the gonads of bass. (Pawson and Pickett, 1996)

tible beneth tin almost transparent ovary wall, and expressed freely with light pressure
VII Spent

Ovary flaccid but not empty, deep red; very thick ovary wall; dense yellow atretic eggs may be visible

Testes flattened and grey, flushed with red or pink, larger than those at stage II or III


Figure B2.1. Logistic maturity ogives (with $95 \%$ confidence intervals) fitted to individual maturity records for sea bass during December-April. Top plot: excluding 2009 data (top); bottom plot: including 2009 data. Points are proportion mature in the raw data. Dotted line is the number of observations per length class.

The parameters of the model $\operatorname{Pmat}(L)=1 /\left(1+\mathrm{e}^{-\mathrm{b}(\mathrm{L}+\mathrm{c})}\right)$ are given below:

|  | Females | MALES |
| :--- | :--- | :--- |
| Intercept (a) | -13.556 | -16.851 |
| Slope (b) | 0.3335 | 0.4861 |
| $\mathrm{c}=\mathrm{a} / \mathrm{b}$ | -40.6488 | -34.6652 |
| L25\% | 37.35 | 32.41 |
| L50\% | 40.65 | 34.67 |
| L75\% | 43.95 | 36.93 |

The logistic model for females and males is:

$$
\begin{array}{rr}
\operatorname{Pmat}(\mathrm{L})=1 /(1+\mathrm{e}-0.3335(\mathrm{~L}-40.6488)) & \text { (females) } \\
\operatorname{Pmat}(\mathrm{L})=1 /(1+\mathrm{e}-0.4861(\mathrm{~L}-34.6652)) & \text { (males) }
\end{array}
$$

The maturation range for females occurs at-ages 4 to 7 , and for males at ages 3-6, as shown by the proportion mature at-age in the same samples used for estimation of length-based maturity ogives (Table B2-2).

Table B2-2. Raw proportion mature at-age in 1982-2003 UK samples from all areas.

|  | FEMALES | 0.00 |
| :--- | :--- | :--- |
| age 2 | 0.00 | 0.27 |
| age 3 | 0.00 | 0.54 |
| age 4 | 0.17 | 0.61 |
| age 5 | 0.21 | 0.98 |
| age 6 | 0.55 | 1.00 |
| age 7 | 0.95 | 0.98 |
| age 8 | 1.00 | 1.00 |
| age 9 | 0.95 |  |
| age 10+ | 1.00 |  |

Data on sea bass maturity have also been collected in the Netherlands since 2005. Methods and data are described by Quirijns and Bierman (2012). For male fish, too few specimens were measured to estimate maturity. Maturity-at-age and length are plotted in Figure B2-2. Note that only few fish were measured in the lowest age and length groups. At age 4,50\% of the females are mature. This is substantially lower than the age at $50 \%$ maturity in the Cefas 1982-2003 samples (Table B2-2), and closer to the ogive from Cefas data including the large 2009 sample (Figure B2-1), for which L50 was around $35 \mathrm{~cm}(-4$ years old). This may confirm that sea bass could now be maturing earlier than in the 1980s-early 2000s, at least for the North Sea. The plot showing ma-turity-at-length for Netherlands samples is not based on enough measurements to show a reliable maturity ogive.


Figure B2-2. Proportion of mature at-age and length (length in m) for female sea bass sampled in the southern North Sea by the Netherlands during 2005 (thick line). The thin line shows the number of fish measured on which the proportion of maturity is based.

B2.4 Larval dispersal, nursery grounds and recruitment
Bass laryae resulting from offshore spawning move steadily inshore towards the coast as they grow and, when they reach a specific developmental stage at around $11-15 \mathrm{~mm}$ in length (at 30-50 days old), it is thought that they respond to an environmental cue and actively swim into estuarine nursery habitats (Jennings and Pawson, 1992). From June onwards, 0 -group bass in excess of 15 mm long are found almost exclusively in creeks, estuaries, backwaters, and shallow bays all along the southeast, south, and west coasts of England and Wales, where they remain through their first and second years, after which they migrate to over-wintering areas in deeper water, returning to the larger estuaries in summer. Several studies indicate the existence of similar bass nursery areas in bays and estuaries on the French coasts of the Channel and Bay of Biscay and southern Ireland.

During the winter, juvenile bass move into deeper channels or into open water, and return in spring to the larger estuaries and shallow bays on the open coast, where they remain for the next 2-3 years.

On the south and west coasts of the UK, juvenile bass emigrate from these nursery areas at around 36 cm TL (age 3-6 years, depending on growth rate), often dispersing well outside the 'home' range, and not necessarily recruiting to their specific parent spawning stock (Pawson et al., 1987; Pickett and Pawson, 2004). It appears that there is substantial mixing of bass at this stage throughout large parts of the populations' distribution range. When they reach four or five years of age their movements become
more wide-ranging and they eventually adopt the adult feeding/spawning migration patterns (Pawson et al., 1994).

Sea temperature has a strong influence on sea bass dynamics, affecting spatial distributions, and also the growth and survival of young bass in nursery areas during the first years of life (Pawson, 1992; WGCSE, 2014).

## B2.5 Natural mortality M

There are no direct estimates of natural mortality available for Northeast Atlantic sea bass. Predation up to around age 4 will be in and near estuaries and bays. As with other fish species it is expected that M will be relatively high at the youngest ages, particularly given the slow growth rate in bass. A variety of methods are given in the literature relating natural mortality rate M to life history parameters such as von Bertalanffy growth parameters k and $\operatorname{Linf}$ (asymptotic length), length or age at $50 \%$ maturity and apparent longevity particularly in an unexploited or very lightly exploited population. The probability of encountering very old bass is partly a function of the interaction of year class strength and sampling rates, as well as mortality, however the occurrence of sea bass to almost 30 years of age suggests low rates of mortality. The observed maximum age of 28 years in sea bass samples in the UK was recorded in the early 1980s, following a period of relatively low fishery landings. Age compositions of recreational fishery caught bass in southern Ireland, presented by stakeholders at IBPNEW 2012, also show ages up to 26 years (Figure B2-3). This stock has been subject to a commercial fishery ban for many year


Figure B2-3. Age composition of bass from samples collected from recreational catches in southern Ireland (data courtesy Ed Fahy, IBP-NEW 2012 meeting).

Inferences on sea bass natural mortality based on some life-history models in the literature are given in IBP-NEW 2012 benchmark assessment section. The inferred values of M , with the exception of the Beverton method, are in the range $0.15-0.22$ (Armstrong, 2012).

A variety of methods are given in the literature relating natural mortality rate M to lifehistory parameters such as von Bertalanffy growth parameters $k$ and Linf (asymptotic length), length or age at $50 \%$ maturity and apparent longevity particularly in an unexploited or very lightly exploited population. The method of Gislason et al. (2010) generates age-varying $M$ values. These methods were applied to the following sea bass life-history parameters by Armstrong (2012):



The inferred values of $M$, with the exception of the Beverton method, are in the range $0.15-0.22$. The average of the Gislason estimates for ages $3-20$ is 0.19 , and the estimates fall below 0.15 by age 11. The value $\mathrm{M}=0.2$ was adopted by WGCSE 2013

## B. 3 Surveys

## B3.1 UK Solent and Thames pre-recruit surveys

The UK has conducted pre-recruit trawl surveys in the Solent and the Thames Estuary since 1981 and 1997 respectively. These surveys all ended in 2009 although the Solent survey was repeated as a one-off survey in autumn 2011 to help provide recruitment indices for the bass benchmark assessment. The location of the surveys and the tow positions are shown in Figure B3-1. Both surveys use a high headline bass trawl, although in the Thames it is deployed as a twin rig and in the Solent as a single rig.


The Solent survey has previously been presented to WGNEW as a combined index across ages in each year class. The index was derived by firstly rescaling the annual mean catch rate per age class to the mean for that age in the survey series, then taking the average of the rescaled values for ages 2-4 in each year class from surveys in MayJuly and September (i.e. up to six values represented in the annual combined index). The Thames survey data were worked up in the same way, although using a different age range for the combined index (ages 0-3). WGNEW 2012 provided the survey data in the more conventional tuning-file format, giving the standardised catch rates (arithmetic mean numbers per 10 minute tow) by year and age, separately for the two surveys (data in assessment report). These surveys have now been discontinued and will not be updated by future working groups unless new resources are allocated.

The spring and autumn Solent survey index series are updated to include the autumn 2013 survey and to amend an error in the autumn survey indices in 2000. The surveys do not show major year-effects, but as noted in previous assessments the autumn (September) survey shows a general increase in recruitment during the 1990s up to the mid2000s, with very low indices for the 2008 onwards year classes, whilst the spring survey shows poor recruitment from around 2002 onwards. Previous Stock Synthesis runs show that the autumn survey is much better fitted than the spring survey. The spring survey is likely to be more strongly affected by weather and by temperature effects on distribution.

The Thames survey series indicates an increase in recruitment from the mid-1990s to early 2000s followed by some poor year classes, possibly a strong 2007 year class, then weak year classes in 2008 and 2009. A problem with the use of the Thames survey is that it may reflect recruitment from spawning that became established in the southern North Sea as the stock expanded in the 1990s under warmer sea conditions, and may therefore not reflect recruitment trends that influence the larger stock components in the Channel and Celtic Sea.

A justification of using the Solent survey as an index of recruitment over the full range of the stock was the results of a statistical, UK-only fleet-based separable model developed by Pawson et al. (2007) and updated by ICES WGNEW (Kupschus et al., 2008). The Pawson et al. model was fitted only using UK age compositions for trawls, midwater trawls, nets and lines, separately for ICES Divisions IVbc, VIId, VIIeh and VIIafg, and was intended mainly to estimate fleet selection patterns. Although it excluded any tuning data, the recruitment-series for each sea area closely resembled the Solent survey indices and to an extent the shorter Thames series, and was able to provide coherent selection patterns by fleet.

The full Solent survey series was subject to a change in gear design in 1993. Some comparative trawling was carried out to develop age-varying calibration factors, but these are poorly documented and the original raw data and calibration results are currently being sought at Cefas. Pending an evaluation of this, the benchmark Stock Synthesis runs included a sensitivity run with the series split into two periods around the gear change. Some additional issues with calibration factors applied to the spring survey were detected during the benchmark, and this is considered later in the sections on model development.

A precision estimate was calculated for the Solent and Thames surveys based on the between-tow variations in catch rate of all the age classes used in the index. For the Solent spring, Solent autumn and Thames surveys, the relative standard errors were $0.42,0.25$ and 0.43 respectively.

IBP-Bass 2014 reviewed the performance of the Solent and Thames surveys and decided to exclude the Solent spring and the Thames survey for the following reasons:

- The Solent spring survey (Table B.3.1) has performed poorly in the assessment, indicating trends in recruitment not in accord with other data. This may be related to temperature or other environmental effects on distribution of small bass in spring. Unusual calibration factors noted in some years in the Solent spring data files require investigation.
- The Thames survey (Table B.3.2) may reflect recruitment from spawning that became established in the southern North Sea as the stock expanded in the 1990 s under warmer sea conditions, and may therefore not reflect recruitment trends that influence the larger stock components in the Channel and Celtic Sea.

Table B.3.1. Time-series of relative abundance indices for seabass age groups 2,3 and 4 from the UK Solent spring and autumn trawl surveys. A change in trawl design took place in 1993.

| May-July |  |  |  | September |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | age 2 | age 3 | age 4 | age 2 | age 3 | age 4 |
| 1981 | 0.00 | 0.30 | 0.25 | No surver |  |  |
| 1982 | 0.51 | 2.17 | 0.16 | 3.25 | 10.10 | 0.38 |
| 1983 | No survey |  |  | 9.87 | 0.91 | 1.88 |
| 1984 | 0.95 | 2.66 | 0.43 | 1.38 | 0.65 | 0.09 |
| 1985 | 0.00 | 10.33 | 2.56 | No survey |  |  |
| 1986 | No survey |  |  | 0.27 |  | 1.31 |
| 1987 | 0.00 | 0.42 | 3.18 | 0.05 | 0.28 | 2.27 |
| 1988 | 0.00 | 0.02 | 0.47 |  |  |  |
| 1989 | No survey |  |  |  |  |  |
| 1990 | 2.84 | 2.48 | 0.00 | 2.81 |  | 0.02 |
| 1991 | 5.78 | 0.62 | 0.09 | 3.08 0.21 0.03 |  |  |
| 1992 | 0.11 | 7.04 | 0.35 |  | , | 0.16 |
| 1993 | 0.05 | 7.33 | 14.02 | $6.65 \sim 3.59$ |  | $4.39$ |
| 1994 | 0.04 | 1.63 | 1.14 | 3.33 | 1.84 | 0.29 |
| 1995 | 0.05 | 1.57 | 0.97 | $\frac{4.83}{5.52}$ | 4.69 | 0.72 |
| 1996 | 1.43 | 4.09 |  |  | 0.43 | 0.11 |
| 1997 | 0.27 | 1.94 |  | 33.62 | 4.52 | 0.06 |
| 1998 | 0.00 | 6.75 |  | 1.22 | 5.50 | 0.61 |
| 1999 | 0.61 |  | 12.30 | 19.37 | 0.67 | 0.87 |
| 2000 | 0.49 | 7. 03 | 1.06 | 6.07 | 11.35 | 0.03 |
| 2001 |  | 6.33 | 3.43 | 34.42 | 3.92 | 1.57 |
| 2002 |  | 1.62 | 0.29 | 7.42 | 3.87 | 0.40 |
| 2003 | 0.06 | 0.32 | 0.38 | 8.37 | 4.60 | 0.59 |
| 2004 | 0.17 <br> 0.05 <br> 0.42 |  | 0.16 | No survey |  |  |
|  |  |  | 0.35 | 13.12 | 7.98 | 0.84 |
| 2006 | 0.44 | 2.47 | 1.03 | 9.51 | 9.21 | 1.02 |
|  |  | 0.50 | 0.50 | 3.42 | 1.78 | 0.30 |
| 2008 | No su |  |  | 18.52 | 6.66 | 0.34 |
| $2009$ | 0.72 | 1.03 | 0.13 | 13.25 | 6.25 | 0.33 |
| 2010 | No survey |  |  | No survey |  |  |
| 2011 | No survey |  |  | 2.25 | 1.39 | 0.42 |
| 2012 | No survey |  |  | No survey |  |  |
| 2013 | No survey |  |  | 1.34 | 0.08 | 0.10 |

Note: September 2000 data amended.

Table B.3.2. Time-series of relative abundance indices for seabass age groups 0-3 from the UK Thames trawl survey.

| Year | AGE 0 | AGE 1 | AGE 2 | AGE 3 |
| :---: | :---: | :---: | :---: | :---: |
| 1997 | 7.737 | 0 | 0.048 | 0.41 |
| 1998 | No sur |  |  |  |
| 1999 | 19.54 | 6.033 | 0.764 | 0 |
| 2000 | 4.015 | 14.74 | 0.832 | 0.089 |
| 2001 | 121.5 | 11.47 | 5.108 | 0.171 |
| 2002 | 469 | 20.71 | 2.716 | 1.093 |
| 2003 | 225.6 | 35.76 | 4.429 | 0.15 |
| 2004 | 238.92 | 44.99 | 7.32 | 1.03 |
| 2005 | 37.04 | 14.49 | 6.86 | 0.75 |
| 2006 | 245.54 | 11.26 |  |  |
| 2007 | No sur |  |  |  |
| 2008 | 107.55 | 50.69 | . 86 |  |
| 2009 | 95.43 | 7.79 | 13.59 | 0.91 |

The UK has undertaken a seine net survey in the Tamar Estuary, since 1985. Additional data are available from the Camel estuary and power stations in the Thames and Severn Estuary. These surveys are used as supporting information and not included in the assessment. Abundance indices for these surveys are given in Table B3.1. The Tamar survey abundance indices need to be updated to include more recent surveys. Seine net surveys in the UK estuaries Fal and Helford also have data on 0-gp and 1-gp bass.

Table B3-1. Abundance indices for 0-gp and 1-gp bass. († discontinued).

|  | Estuary seine surveys |  | Power station screen |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | Tamar (0-group) | Tamar (1-group) | Camel | Severn | Thames <br> IVc |
|  | VIIe | VIIe | VIIf | VIIf |  |
|  |  |  |  | 3 |  |
| 1973 |  |  |  | 4 |  |
| 1974 |  |  |  | 1 |  |
| 1975 |  |  | 15 |  | 78 |
| 1976 |  |  | 127 |  |  |
| 1977 |  |  |  | - |  |
| 1978 |  |  |  | - | 5 |
| 1979 |  |  |  |  |  |
| 1980 |  |  |  |  |  |
| 1981 |  |  |  | 16 | 21 |
| 1982 |  |  | 12 |  | 56 |
| 1983 |  |  | 30 | 226 | 83 |
| 1984 |  |  | 134 | 8 | 62 |
| 1985 | 0.663 | 0.385 | 22 | 11 | 76 |
| 1986 | 0.005 | $0.014$ |  | 3 | 14 |
| 1987 | 0.032 | 0.062 |  | 96 | 116 |
| 1988 | 1.484 | 1.284 | 48 | 98 | 54 |
| 1989 | 2.348 | 2.389 | 112 | 446 | 610 |
| 1990 | 1.038 | 1.516 | 89 | 25 | 433 |
| 1991 | 0.076 |  | 50 | 300 | 64 |
| 1992 | 216 | 2.431 | 25 | 280 | 104 |
| 1993 |  | 0.913 | 22 | 202 | 131 |
|  | 126 | 0.346 | 134 | - | 26 |
|  | 2.356 | 1.294 | - | - | 27 |
| 1996 | $0.102$ | 0.047 | 119 | 242 | + |
|  | $119$ | 1.299 | 102 | + |  |
| 1998 | 2.082 | 3.170 | 264 |  |  |
| 19 | 1.215 | 0.937 | 56 |  |  |
| 2000 | 0.340 | 1.185 | 133 |  |  |
|  | 0.351 | 0.129 | + |  |  |
| 2002 | 2.098 | 3.179 |  |  |  |
| 2003 | 0.965 | 1.067 |  |  |  |
| 2004 | 1.453 | 0.261 |  |  |  |
| 2005 | 0.522 | 0.169 |  |  |  |
| 2006 | 0.186 | 0.203 |  |  |  |
| 2007 | 0.475 | 1.308 |  |  |  |
| 2008 | 1.275 | 1.229 |  |  |  |
| 2009 | 0.460 |  |  |  |  |

## B3.3 Evhoe survey: France

Sea bass are caught in small numbers in the French Evhoe trawl survey, which extends to the shelf edge in Subareas VII and VIII but also extends into coastal areas of the Bay of Biscay and the Celtic Sea where bass may be caught (cf the station map). Less than $10 \%$ of the stations have bass catches in most years. A mean of 0.5 sea bass per trawl has been recorded from 1987. Abundance indices are calculated as stratified means.


Figure B 3-2. Station positions for French Evhoe bottom-trawl survey (not used in assessment).

## B3.4 Channel Ground Fish Survey (CGFS): France

Raw data on sea bass from the French scientific trawl survey "Channel Ground Fish Survey - CGFS" were not available for the previous benchmark in 2012 (IBPNEW, ICES, 2012). Details of the survey are given in Coppin et al. (2002), which includes a full description of the GOV traw used in October each year at the 82 stations in ICES Division VIId shown in Figure B3.3. The majority of sea bass are caught in the coastal waters of England and France (Figure B3.3). The abundance indices from all the stations give similar trends as from a subset of stations in the main coastal areas, and trial runs with SS3 gave similar trends. Therefore, for further SS3 development, the indices calculated from all the area are used.
abundance indices are calculated applying a stratified random swept-area based estimator. Strata correspond to ICES statistical rectangles. Swept-area is calculated using wingspread. As this is a stratified swept-area based indicator, uncertainty is based on between haul variance within a strata and summation of variances across strata. Full methodology is presented in the WD_01, available in Annex 2 of the IBPBass report 2014. The trends in both the index and in the proportion of stations with sea bass show some similarities to the trend in total biomass estimates from the ICES WGCSE 2013 update assessment using Stock Synthesis, which lent a priori support to the use of the index in the assessment. The swept-area indices of abundance, the percentage of stations with sea bass, and the variance of the estimates are included in the WGCSE 2014 report and will be updated annually. The length composition of the survey index is calculated and is also input to Stock Synthesis.

The precision of the swept-area indices appears unrealistically high in some years (e.g. 0.025 in 1991), which may indicate that the index trends are driven largely by the incidence of positive catches. Modelling of the data using delta lognormal models may provide more realistic precision. During trial Stock Synthesis runs, the use of the CVs resulted in an inability to fit the selection curve for the survey due to individual years being given far too much weight. Relaxing the CVs to 0.30 for all years except the first three years (set to 0.6 given the very low incidence of positive stations) allowed the model to fit the length compositions more closely over the series. The annual indices are therefore input to Stock Synthesis with a CV of 0.30 for all years. The effective sample sizes for the annual survey length composition data are set at the number of stations with sea bass length data. The length compositions for the first three years (1988-1990) are excluded from the assessment due to very small sample sizes although the aggregate indices are retained.


Figure B.3.3. Left: stations fished during the Channel Groundfish Survey carried out annually by France. Right: distribution of total catches of sea bass over the survey series.

## B. 4 Commercial Ip

ICES IBPNew2012 evaluated a range of commercial fishery lpue series for French and UK fleets operating in Areas IV and VII. The UK analysis of official catch statistics involved filtering individual trip data to include only trips in ICES rectangles where bass catches have been recorded historically. UK vessels of 10 m and under, for which historical landings data are very uncertain, were found to have a wide range of lpue trends depending on gear and area fished, often showing a very steep increase since the mid2000s (Armstrong and Maxwell, 2012). This may be partly a consequence of more accurate reporting caused by the Registration of Buyers and Sellers regulations after its introduction in 2005, but may also represent a bias caused by increased targeting of sea bass by vessels with insufficient quotas for other stocks or trying to develop track record.

With some exceptions (e.g. trawlers in VIId), UK $>10 \mathrm{~m}$ vessels tended to show different lpue trends to 10 m and under vessels. Relative trends of seabass lpue for $70-99 \mathrm{~mm}$ mesh UK otter trawls (1985-2011) and French otter trawlers (2000-2010) operating in IVbc,VIId, VIIeh and VIIafg showed a general trend of increase in the 1980s and 1990s, followed by a levelling off and a decline after 2009 (Figure 10.1.2.7, from WGCSE 2013).

The trends for $>10 \mathrm{~m}$ UK and French trawlers in IV\&VIId and in VIIe closely matched the trend in total stock biomass estimates from the final WGCSE 2013 Stock Synthesis assessment whereas the UK trawlers in VIIa,f,\&g had a much lower lpuein the early part of the time-series. These results indicated a potential for development of fishery lpue series for inclusion in future development of SS3 for sea bass, using more sophisticated trip filtering and using more statistical approaches such as delta-lognormal modelling with GLMs to develop standardised series. However, IBP-Bass focused more on developing the survey data inputs and did not progress the development of the fishery lpue series which remains a task for the future.

## B4.1 UK bass logbook scheme

The UK bass logbook scheme is described in Section B1.1. Although the survey has severe limitations for estimation of total bass landings for UK vessels, individual logbooks provide time-series of varying duration on catch-rates of individual vessels using specific gears. The logbooks with sufficient data cover eight gear types within trawls, nets and lines, covering mainly 10 m and under vessels, excluding recreational vessels. The total numbers of logbooks have declined from 50-60 in earlier years to below 20 in recent years. No logbooks were issued in 2008:

| Year |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Region | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2009 | 2010 |
| 1 | 16 | 19 | 14 | 12 | 13 | 8 | 6 | 0 | 3 | 3 |
| 2 | 15 | 15 | 13 | 14 | 7 | 10 | 5 | 0 | 3 | 2 |
| 3 | 2 | 5 | 3 | 5 | 5 | 5 | 7 | 0 | 3 | 3 |
| 4 | 4 | 5 | 6 | 7 | 1 | 3 | 4 | 0 | 3 | 1 |
| 5 | 9 | 10 | 9 | 4 | 2 | 5 | 6 | 2 | 1 | 1 |

(Region 1: North Sea IVbc, 2: eastern Channel VIId; 3: western Channel VIIeh; 4: Celtic Sea (VIIfg); Irish Sea (VIIa). The trend in number of records per year shows roughly the same pattern across gears:
An exploratory GAM method was developed (Armstrong and Maxwell, 2012) to extract a common temporal trend in lpue from the individual series for ICES Areas IVbc\&VIId, VIIeh and VIIafg (referred in the models as areas $1 \& 2,8$ and $4 \& 5$ ). This is analogous to combining series of tree ring counts from timbers of various ages to give a single series describing climate changes. The general method involves estimating logbook factors and year factors (and interactions) to minimise residual model error. Following initial model development and evaluation, a negative binomial error distribution with log link was selected. This can accommodate zero values and allows for the variance to increase with the mean. Working with a log link implies that the estimated trend with year is multiplicative not additive. The R command showing the exact options used for areas $1 \& 2$ combined (North Sea and VIId) is:
bass.gam3.12 = gam(lpue $\sim$ factor(BookGear) +s (Year, $\mathrm{k}=10$, $\mathrm{bs=}=$ "ts"), family=neg$\operatorname{bin}(c(1,10))$, optimize $=$ "perf", data=bass.dat, subset=ARegion=="1and2")

Fitted trends and confidence intervals suggest an increasing lpue trend in regions $1 \& 2$ (North Sea \& VIId), and 3 (VIIeh) (Figure B4-1). A relatively flat trend and possible recent decline is indicated in regions $4 \& 5$ (VIIafg) although the recent trend is highly imprecise. Residual checks indicate the model assumptions are reasonable. Model diagnostics and sensitivity to smoothing and other parameters are given in Armstrong and Maxwell (2012).


Figure B4-1. Cefas bass logbook lpue: Selected model for combined regions, plots showing year effects from a fitted model with separate mean value for each book number-gear combination and negative binomial error distribution, dashed lines are a $95 \%$ confidence interval.

## B4.2 UK fleet lpue based on official catch dataseries

Armstrong and Maxwell (2012) review trends in UK commercial fishery lpue for sea bass in the North Sea (IV), eastern Channel (VIId), western Channel (VIIe) and Irish/Celtic Seas (VIIafg) from 1985-2011, and evaluate the possibility of using the timeseries as relative abundance estimates fortuning stock assessment models.
Gears which catch bass are targeted at a variety of species, and the fishing effort is distributed across many areas where sea bass have zero or very low probability of capture. A number of approaches are possible to subset fishing trips to include only those that have a probability of catching the species for which lpue is to be estimated. One approach (Stephens andMacCall, 2004) is to cluster fishing trips according to species that occur in association, and use only the cluster with the species on interest for estimating lpue. This method has not yet been applied to UK data. An alternative method to subset trips was applied. This involved (a) selecting gear types that account for $\sim 95 \%$ of the total bass landings in each area since 2005; (b) for the selected gears and areas, identify ICES rectangles accounting for $\sim 95 \%$ of the total bass landings since 1985. Annual lpue was then estimated for each area and gear, separately for vessels of 10 m (LQA) and under and $>10 \mathrm{~m}$ vessels. The LOA split is important because reporting of landings and effort of 10 m and under vessels has been very uncertain historically, particularly prior to the introduction of Buyers and Sellers regulations in 2005. Lpue of 10 m and under vessels may be very inaccurate prior to 1995.
It was not possible to evaluate the effect of any increase in targeting of bass by individual vessels using the selected gear types in the selected rectangles, or effects of technology creep. Increased targeting is likely to have happened for vessels with increasingly limited quotas for other species such as cod and which have switched to non-TAC species such as sea bass. For some gears, such as beam trawls, sea bass are not targeted and are purely a bycatch.

Too many lpue series have been examined to reproduce in the Stock Annex, but can be viewed in Armstrong and Maxwell, 2012.

## B4.2 French Ipue sets

Lpue of French trawlers in IVb,c, VIId and VIIeh is available from 2000 when Ifremer has estimated landings by ICES Divisions. A recent study has developed indices as $\mathrm{kg} /$ per day based on data from auction's sales. This study was carried out on French bottom trawlers (less than 18 m ), having a fishing strategy with the least distant random sampling; this fleet usually doesn't target sea bass. Large bias can be caused where: 1. an auction sale corresponds to several days of fishing, 2. technological advances are not taken into account, and 3. changes in fishermen's strategies are not taken into account. Never the less, for information, those from the Channel and North Sea have been compared to the UK Otter trawls lpue, and similarities shown on Figure B42 are observed.


Figure B4-2. UK fleet lpue based on official catch dataseries, compared to the French lpue sets based on auction hall sales.

In 2015 for WGCSE, a study "French Logbook data analysis 2000-2013: possible contribution to the discussion of the sea bass stock(s) structure/annual abundance indices. Alain Laurec, M.Drogou"has been conducted and presented in a Working Document (WD_seabass_France_Abundance_Index.pdf: ICES, WGCSE 2015).

Daily catch rates per vessel, grouped within months and ICES squares, have been anysed basically through a multiplicative two factors model. The two factors, namely the fishing vessel effect and the stratum effect. A stratum corresponds to a so-called ICES square, a month and a year.

First conclusions provide a basic hypothesis about stock structures and spawning migrations, and directly related to this discussion apparent abundance index have been produced covering various option/areas.

The preliminary results of the study are considered promising by the group. Even if it's still under-development and should be benchmarked to use it directly in the assessment, some comparison of the Index from two various option with the SSB is presented in Figure B4-3. This shows similar trends in stock perception conforting results of the assessment (but question on the degree of the trends).


Figure B.4.3. Bass-47: Trends in commercial lpue index for French fleets overlaid on this year's update assessment estimates of spawning-stock biomass ( $+/-2$ standard errors). Top: index based on data from all twelve months; bottom: index excluding fishing trips during spring spawning season.

## B.5. Other relevant data

None.

## C. Assessment: data and method

## Software used and model options chosen

Model used: Stock Synthesis 3 (SS3) (Methot, 2010)
Software used: Stock synthesis v3.23b (Methot, 2011)
WGCSE 2013 conducted an update assessment using Stock Synthesis 3 (SS3) (Methot, 2010). The software used was Stock synthesis v3.23b (Methot, 2011), according to the

Stock Annex developed by ICES IBPNEW 2012 with inclusion of fishery data for 2012. The assessment requires a modelling framework capable of handling a mixture of age and length data for fisheries and surveys (fleet-based landings; landings age or length compositions, age-based survey indices for young bass) and biological information on growth rates and maturity. Landings-at-age were available for four UK fleets from 1985 onwards, whereas French fleets had length composition data that were available only since the 2000s. The Stock Synthesis assessment model was chosen, primarily for its highly flexible statistical model framework allowing the building of simple to complex models using a mix of data compositions available. The model is written in ADMB (www.admb-project.org), is forward simulating and available at the NOAA toolbox: http://nft.nefsc.noaa.gov/SS3.html.

A mixed age-length model was fitted by WGCSE 2013 as the base case, with a lengthonly model for comparison. Some adjustments were made by WGCSE 2013 to the model: i) UK fishery compositions for 2012 were input to the age-length model as length compositions because age compositions were not available; ii) the UK midwater trawl series was reduced to 1996 onwards and was input as length compositions because unusual length-based selection curve parameters were obtained when inputting the data as age compositions; iii) recruit deviations were estimated back to 1965

IBPBass (ICES 2014) addressed the following recommendations of WGCSE 2013 for developing the assessment during the inter-benchmark meeting. Work completed is indicated in parenthesis:

- Source and review information on historical catches and develop plausible scenarios including over the $20+$ year burn-in period for the assessment [some investigations were pursued in France but yielded no clear information on pre-1985 landings]
- Review the derivation and quality of historical fishery length/age composition data not done beyond the information on sampling intensity and coverage already available];
- Expand UK fishery age compositions to all true ages [done, see below];
- Rationalise the fleet definitions, and reduce to the minimum sufficient to provide robust SS3 stock trends [done, see below]; fishing mortality on older ages [fishery-dependent abundance indices were not considered other than some information presented in Section 2 on lpue of fleets in the Netherlands];
- Collate and evaluate other survey data on bass abundance that could be incorporated in the model [French Channel Groundfish Survey was evaluated and incorporated in the assessment];
- Determine the most robust approach to incorporating mean length-at-age and length-at-age distributions in SS3 [Not done];
- Investigate potential biases in using combined-sex growth parameters [Not done];
- Further explore the sensitivity of the assessment to decisions on model structure and inputs [See model development and sensitivity analyses carried out below];
- Consider if simpler assessment approaches are warranted [IBPbass focused exclusively on Stock Synthesis to try and make best use of all available data].


## Deviations from Stock Annex

The Stock Annex has been revised to reflect the approaches developed by IBP-Bass. The update assessment follows exactly the IBP-Bass data inputs and methods.

The assessment requires a modelling framework capable of handling a mixture of age and length data for fisheries, including data for French fleets that have length composition data but no age composition data, and for which the length data were available only since the 2000s. The Stock Synthesis (SS) assessment model was chosen, primarily for its highly flexible statistical model framework allowing the building of simple to complex models using a mix of data compositions available. The model is written in ADMB (www.admb-project.org), is forward simulating and available at the NOAA toolbox: http://nft.nefsc.noaa.gov/SS3.html. For European sea bass a model was built using Stock Synthesis 3 (SS3) version $3.29 b$ to integrate the mix offisheries and survey data available (fleet-based landings; landings age or length compositions,ages-based survey indices for young bass and length-based CGFS survey indices) and biological information on growth rates and maturity.
Developments of the Stock Synthesis model by IBP-Bass including the following improvements:

- Increased the plus-group from $12+$ to
- Combined UK trawl, net and line fleets into a single fleet with double-normal selectivity.
- Fitted age-based selectivity to all UK commercial fleets.
- Included the Channel Groundfish survey with length-based double-normal selectivity.
- Excluded the Thames and Solent spring juvenile surveys.
- Included a fixed vector of recreational fishing mortality at age, consistent with recreational fishery removals of around 1500t in 2012.
- Reduced the natural mortality rate from 0.20 to 0.15 .

Full details of the reasons for these changes, and diagnostics of model fits with these changes included, can be yiewed in the IBP-Bass report. Information is given below on some of the key decisions of IBP-Bass concerning the final structure of the assessment model.

## Changes to the plus-group

UK fishery age composition data show year-class structure extending well beyond the revious oldest true age used in the assessment (eleven years). This includes year classes well before the first year of input data (see Figure 10.1.2.1). IBP-Bass carried out runs with plus groups set at $12+, 16+, 18+$ and $20+$. The $16+$ setting appeared sufficient to improve estimation of early recruit deviations prior to 1985, without causing problems of zero catch estimates appearing in the data file at old ages. All further runs were conducted with $16+$ group.

## Selectivity patterns for UK commercial fleets

The WGCSE 2013 assessment treated all commercial fleets from the UK and France as having asymptotic length-based selectivity. This reduced the number of parameters per fleet to only two, an important consideration when four UK fleets and one French fleet were being modelled. Having collapsed the UK trawls, nets and lines to a single fleet, IBPBass explored the use of a double normal selectivity pattern, which appeared
to be appropriate for trawls and nets from the results of the Pawson et al. (2007) and Kupschus et al. (2008) separable model applied to separate UK fleets.

The form of the selectivity pattern was investigated initially using a simple cohort analysis applied to the aggregated catch-at-age for the four UK fleets since 2013. The terminal F for this aggregate "pseudo cohort" was adjusted until the pattern of partial F's for the UK midwater trawl and lines fleets were as close as possible to asymptotic, as these fleets target a wide range of adult and juvenile bass in inshore and offshore waters. The pattern of partial F's for the trawls and nets fleets were then revealed as being strongly domed, confirming previous results in the Pawson et al. (2007) separable model (Figure 10.1.3.1). Input initial parameters and bounds for a double-normal selectivity function (a selectivity form recommended in the Stock Synthesis manual) were derived for UK trawls, nets and lines using the spreadsheet developed for Stock Synthesis. The fitted selectivity for this fleet and for the midwater trawl fleet in the update Stock Synthesis run closely match the selectivity patterns given by the empirical approach using cohort analysis (Figure 10.1.3.1).

The WGCSE 2014 update assessment uses the same selectivity models and input selectivity parameters as agreed by IBP-Bass, fitting the parameters using soft bounds rather than priors with hard bounds. This resulted in fitting ten selectivity parameters for commercial fleets (six for UK trawls, nets and lines, and two each for midwater trawls and combined French fleets. The "other fleet" was assumed to have the same selectivity as the French fleet).

## Natural mortality and stock-recruit steepness

The value (or vector) of natural mortality used in an analytical assessment is a key parameter determining the estimated productivity, abundance and MSY or other reference points (RPs). The assumed shape of the stock-recruit curve acts with the assumed M to constrain any possible biological reference points (Mangel et al. 2013). A key parameter defining a stock-recruit curve is steepness, defined as the ratio of recruitment from an unfished population to recruitment when the spawning-stock biomass is at $20 \%$ of the unfished level.

IBP-Bass explored the performance of the Stock Synthesis model for a range of different values for natural mortality and stock-recruit steepness, using a similar model formulation to the WGCSE 2014 update but excluding the recreational fishing mortality vector. The total of negative log likelihood was compiled for the range of combinations, along with the SSB depletion in 2013 from the virgin SSB, and the relative standard error of theSSB in that year.

Otal negative log likelihood tended to be lowest at steepness values approaching unity, with the greatest tendency at low values of $M$, and likelihood also decreased with increasing M (Figure 10.1.3.3. top left). Despite the lower value of negative log likelihood, the relative standard errors of the SSB estimates for 2012 (from the inverse Hessian) increased with increasing $M$ but were almost unaffected by steepness (RSE values at steepness 0.999 were 0.158 at $\mathrm{M}=0.15 ; 0.164$ at $\mathrm{M}=0.20 ; 0.175$ at $\mathrm{M}=0.25$ and 0.199 at $\mathrm{M}=0.30$ ). The unusual value at steepness $0.8, \mathrm{M}=0.3$ in Figure 10.1.3.3 was a result of the selectivity curve for the combined UK trawls, nets and lines flipping from a domed to an asymptotic pattern. Otherwise, the values show smooth relationships with input M and steepness.

The depletion of SSB in 2013 compared with the virgin SSB was progressively lower as M was increased (Figure 10.1.3.3. top right), but was far less sensitive to steepness.

Recruitment in sea bass has varied widely in response to environmental factors including conditions in the estuarine and other inshore nursery habitats. There is almost no information to indicate declining recruitment at lower SSB and to discern the true value of steepness. A wide range of values appears plausible (Figure 10.1.3.3. lower plots), though the model fit slightly favours the (biologically implausible) steepness value of unity.
IBP-Bass decided to retain a fixed steepness value of 0.999 given the relative insensitivity of the assessment to this parameter. This means that F or biomass at MSY cannot be estimated, and that proxy $\mathrm{F}_{\mathrm{mSy}}$ reference points have to be specified using yield-perrecruit and spawner biomass per recruit computations.
The tendency for likelihoods to improve with increasing $M$, despite the inferences that a relatively low M is consistent with sea bass life-history traits and maximum age approaching 30 years, however suggested that additional mortality associated with the recreational fishery could perhaps be accommodated within the model.

## Incorporating information on recreational fishery catches

Recent surveys in France, England and the Netherlands have proyided estimates of recreational fishery removals that are around a third of the reported commercial fishery landings in each country (see Section 2 ). Taking possible hooking mortality of released fish into account, the total recreational removals from the stock may be as high as 1500 t . The relative standard error of this estimate is likely to be in the range $0.2-0.3$ based on the information available on the component surveys. With only one combined survey estimate available, IBPBass could see no easy way to incorporate this directly into Stock Synthesis.
However, by not representing recreational fishery removals in the assessment, the estimates of total mortality derived from the commercial fishery age and length composition data are attributed only to natural mortality and to commercial fishing mortality. In reality, part of the observed total mortality is attributable to the unaccounted-for recreational fishery removals. This becomes a problem when forecasting stock size and yields based on multipliers applied to the apparent commercial fishing mortality, when in fact only part of the apparent $F$ is due to the commercial fleet.

IBP-Bass explored an approach for evaluating the magnitude of recreational fishery F by including a fixed vector of recreational F-at-age in the assessment, added to the fixed natural mortality vector. The recreational F vector, with a plausible selectivity pattern for the harvest (kept component), was then scaled iteratively until the expected recreational removals in 2012 were around 1500 t ; the recent total recreational removals estimate from the surveys in France, England and the Netherlands. This is equivalent to treating recreational fishers as a predator whose historical abundance and predation is largely unknown, but are assumed to impose a constant mortality in exactly the same was as all other predators are subsumed into a constant $M$ vector.

There is some evidence from a series of surveys in England and Wales since 1970 that the number of people going sea angling has fluctuated without obvious increasing or decreasing trend (Table 10.1.3.1). Part of the observed variability will relate to differences in survey methodology, but all are based on some form of sampling of the population as a whole. Sea bass has been a prized target for recreational sea anglers in England and Wales (and southern Ireland) over a much longer period than the current assessment, and sea bass angling was developed to a high level of technical skill and knowledge of the species as far back as the 1970s. There is no information on the actual effort expended by the angling population on sea bass as the stock has changed in
abundance, or on changes in efficiency, but an assumption of a constant recreational fishing mortality is a reasonable first approximation for evaluating recreational F.

## Recreational fishery selectivity

It is first necessary to identify an appropriate selectivity function to characterise the selectivity of recreational fishing. The recreational fishery in England and Wales is predominantly rod-and-line, but it is known that other gears (especially fixed or driftnets, or seines) are deployed recreationally throughout Europe. However, in France around $80 \%$ of the recreational fishing catch is from sea angling.

The length frequency of retained sea bass recorded during the recent Sea Angling 2012 survey in England (Anon, 2014) is very noisy due to small samples (many bass are released) and a clear tendency in some cases for lengths to have been reported to the nearest 5 cm (Figure 10.1.3.2). However, the distribution is clearly much more similar to that of the UK commercial line fishery than to the other commercial fishery gear groups. It was therefore decided to use the UK commercial line fishery selectivity (agebased) to represent the combined selectivity and retention of the recreational fishery for the whole of the bass stock. The line fishery selectivity-at-age was obtained from a Stock Synthesis run treating lines as a separate fleet with asymptotic selectivity estimated from UK age composition inputs. The assumption regarding selectivity may not hold beyond the UK, but was adopted in the absence of other information. Weights-atage in the retained catch were assumed to be the same as in the UK commercial fishery landings given the strong influence of minimum landing size of 36 cm on retention in both fisheries.

## Natural mortality component of combined mortality vector

A range of natural mortality values from 0.15 to over 0.20 can be inferred from lifehistory parameters and observed maximum age (see earlier). IBPNew 2012 adopted a value of 0.20 for all ages. The SS3 runs carried out by IBP-Bass at different values of M indicated that likelihoods improved as M was increased, but the precision of recent biomass estimates declined, and at $\mathrm{M}=0.30$ some instability in fitting selectivity parameters was evident. IBP-Bass therefore set up a run using the lower bound of M inferences $(M=0.15)$ in order to avoid the combined $M$ and recreational $F$ going too high, and scaled the recreational fishery selectivity so that the calculated recreational landings in 2012 were 1500t. The selectivity and recreational F vector achieving this are given in Table 10.1.3.2. The recreational $\mathrm{F}(5-11)$ is 0.09 derived from a multiplier of 1.0 applied to the selectivity curve. The combination of $\mathrm{M}=0.15$ and recreational F leads to a combined mortality vector increasing from 0.15 at the youngest ages to 0.25 at-age 8 with a combined $\mathrm{F}(5-11)$ of 0.24 , compared with the total M value of 0.20 used by WGCSE 2013. The sensitivity of the assessment to different absolute values of recreational fishery harvest in 2012, and to the choice of $\mathrm{M}=0.15$ or 0.20 , is explored later in this report

The structure and input data/ parameters of the IBPbass (2014) model, used for the update assessment by WGCSE 2014, are summarized below:

The structure and input data/ parameters of the IBPbass revised SS3 model are summarized below:

## Model structure

- Temporal unit: annual based data (landings, survey indices, age frequency and length frequency);
- Spatial structure: One area;
- Sex: Both sexes combined.


## Fleet definition

Four fleets defined: 1. UK bottom trawls, nets and lines; 2. UK midwater pair trawls; 3. French fleets (combined); 6. Other (other countries and other UK fleets combined). [WGCSE 2013 assessment modelled selectivity separately for UK trawls, midwater trawl, nets and lines].

## Landed catches

Annual landings in tonnes from 1985 to final year for the four fleets from ICES Subdivisions IVb and c, VIIa, d-h. French data were as provided by Ifremer.

## Abundance indices

Channel Groundfish Survey in VIId in autumn (France), 1988 to present: total sweptarea abundance index and associated length composition data. Number of stations with sea bass is used as input effective sample size. Input CV for survey $=0.30$ all years. First three years of composition data are excladed. [Survey not included in WGCSE 2013 assessment].

Cefas Solent survey in autumn (VIId). Years 1986 to 2011; 2013. Three independent abundance index series were defined, each being a single age group ( $2,3,4$ years old). They are treated as three independent surveys (following a recommendation from R . Methot) to circumvent difficulties in estimating selectivity parameters for a survey series comprising only three young age groups, although this approach loses covariance information due to year-effects in the survey. [Solent spring and Thames survey included in WGCSE 2013 assessment].

Fishery landings age composition data: UK fleets
Age bins: 0 to 15 with a plus group for ages 16 and over. Age compositions for UK fleets are expressed as fleet-raised numbers-at-age, although they are treated as relative compositions inSS3. Year range for UK trawls/nets/lines: 1985 to present; UK midwater pair trawl:1996 to present.

## Length composition data: French fleets

The length bin was set from 4 to 100 cm by 2 cm intervals. Length compositions for the following fishing fleets were used: French all fleets combined: 2000 to present.

## Model assumptions and parameters

The following table summarises key model assumptions and parameters. Other parameter values and input data characteristics are defined in the SS3 control file BassIVVII.ctl, the forecast file Forecast.SS and the data file BassIVVII.dat as used by IBPBass 2014 and by WGCSE 2013. Model inputs and year ranges (for WGCSE 2015) are in Figure C1.

Key model assumptions and parameters from the WGCSE 2014 update assessment.

| Characteristic | Settings |
| :---: | :---: |
| Starting year | 1985 |
| Ending year | 2013 |
| Equilibrium catch for starting year | 0.82* landings in 1985 by fleet. |
| Number of areas | 1 |
| Number of seasons | 1 |
| Number of fishing fleets | 4 |
| Number of surveys | two surveys: CGFS; Solent autumn survey (Solent spring and Thames survey removed). |
| Individual growth | von Bertalanffy, parameters fixed, combined sex |
| Number of active parameters | $68$ |
| Population characteristics |  |
| Maximum age | 30 |
| Genders |  |
| Population length bins | $4-100,2 \mathrm{~cm}$ bins |
| Ages for summary total biomass | 0-30 |
| Data characteristics |  |
| Data length bins (for length structured fleets) | 14-94,2 cmbins |
| Data age bins (for age structured fleets) | 0-16 |
| Minimum age for growth model |  |
| Maximum age for growth model | 30 |
| Maturity | Logistic 2-parameter - females; L50 $=40.65 \mathrm{~cm}$ |
| Fishery characteristics |  |
| Fishery timin | -1 (whole year) |
| Fishing mortality method | Hybrid |
| Maximum F | 2.9 |
| Fleet 1: UK Trawl/nets/lines selectivity | Double normal, age-based |
| Fleet 2: UK Midwater trawl selectivity | Asymptotic, age-based |
| Fleet 3: Combined French fleet selectivity | Asymptotic, length-based |
| Fleet 4: Other fleets/gears selectivity | Asymptotic: mirrors French fleet |
| Year-invariant recreational fishing mortality vector $(F(5-11)=0.09)$ | Asymptotic, age-based (fixed, not estimated). Added to M vector |
| Survey characteristics |  |
| Solent autumn survey timing (yr) | 0.83 |
| CGFS survey timing (yr) | 0.75 |
| Catchabilities (all surveys) | Analytical solution |
| Survey selectivities: Solent autumn: | [all survey data entered as single ages; sel = 1] |
| Survey selectivities: CGFS | Double normal, length based |
| Fixed biological characteristics |  |
| Natural mortality | 0.15 |
| Beverton-Holt steepness | 0.999 |
| Recruitment variability ( $\sigma$ ) | 0.9 |
| Weight-length coefficient | 0.00001296 |
| Weight-length exponent | 2.969 |


| CHARACTERISTIC | SETTINGS |
| :--- | :--- |
| Maturity inflection (L50\%) | 40.649 cm |
| Maturity slope | -0.33349 |
| Length-at-age Amin | 19.6 cm at Amin $=2^{1}$ |
| Length-at-Amax | 80.26 cm |
| von Bertalanffy k | 0.09699 |
| von Bertalanffy Linf | 84.55 cm |
| von Bertalanffy t0 | -0.730 yr |
| Std. Deviation length-at-age (cm) | SD $=0.1166^{*}$ age +3.5609 |
| Age error matrix | $\mathrm{CV} \mathrm{12} \mathrm{\%} \mathrm{at-age}$ |
| Other model settings |  |
| First year for main recruitment deviations for <br> burn-in period | 1965 |
| Last year for recruit deviations | 2011 (last year class with survey indiees) |

${ }^{1}$ as recommended by R. Methot after scrutinizing earlier SS3 runs during IBPNEW 2012, and used by IBPNEW and WGCSE. The WGCSE 2013 tabulated the original value of 5.78 cm at-age 0 in error.


## C. 2 Assessment procedure

The model is run with the executable file SS3.exe in the same folder as the following files:

| BassIVVII.ctl | SS3 configuration file |
| :--- | :--- |
| BassIVVII.dat | SS3 data inputs |
| Starter.SS | SS3 startup file |
| Forecast.SS | SS3 forecast file |

Results are ouput in the same folder (key results file is "results.sso"). Plots can be generated using r4ss after calling library(r4ss), using the following code (adjusted with correct path name):

```
age <- SS_output(dir= 'C:/Users/ma02/Documents/ICES/WGCSE2014/Bass47/SS3
update assessment')
SS_plots(replist=age,pdf=F,png=T,dir='C:/Users/ma02/Docu-
ments/ICES/WGCSE2014/Bass47/SS3update assessment')#,uncertainty=F)
```

Retrospective analysis is done with the output files from the base run in the same folder as the file retro.bat. For five retrospectives, six Starter files are included. The base file Starter.SS includes the following code nine lines from the bottom:
-5 \# retrospective year relative to end year (e.g. -4)

The five retrospective Starter files use the name convention Starter-5; Starter-4; Starter3; Starter-2; Starter-1, amending the command -5 \# retrospective year relative to end year (e.g. -4) to reflect the year peel stated in the file name. A piece of code "RetroPlots_R4SS" is available to plot the retrospectives although an Excel file is currently used to read the results from each of the Report.sso files imported into worksheets.

For the WGCSE 2014 assessment, a year-invariant recreational fishing mortality vector $(F(5-11)=0.09)$ was added to the annual M of 0.15 , specified as age-specific values in


The recreational component was arrived at iteratively to generate a total recreational fishery landing close to 1500 t for 2012.

Future runs may need to revise this to generate the same recreational catch.
When the end year for the Stock Synthesis run is specified as the last year with fishery data, the Report.sso file contains estimates of biomass and numbers only to the start of the final year with data, and Zs only to the year before the final one. A work-around to get biomass and numbers for survivors at the end of the last year with data, and Zs for the final year with data, the end year can be specified as the year after the last with
data. F values, as used by ICES, are not generated automatically by Stock Synthesis but can be computed from the Zs after subtracting M .

## D. Forecast

Due to the additional complexity of adding a fixed recreational fishing mortality vector for removals (harvest), and the time required to configure Stock Synthesis to mirror the ICES procedures for short-term forecasts, IBPBass decided not to try and develop a forecast procedure within Stock Synthesis for use by WGCSE. This unfortunately loses the ability to provide MCMC confidence intervals around the assessment and forecasted variables, and the forecasts are entirely deterministic. Management options involving biological reference points (BRPs) adopt BRPs conditional on the assumptions in the assessment regarding M , selectivity, maturity, weights-at-age, etc. The procedures for deriving inputs for the short-term forecast are described below.

## D. 1 Estimating year-class abundance

Stock Synthesis should be set up to estimate recruit deviations only until the last year that has survey data to make the estimates. E.g. including the Solent survey indices for ages 2-4 in 2013 means that the last year class tuned by a survey index is 2011 (two year-olds in 2013). SS3 will put a value from the fitted stock-recruit curve for later year classes. WGCSE overwrites these later year classes using the long-term (1985 onwards) geometric mean, or a short-term GM if there is a persistent reduction in recent recruitment. The numbers-at-age for the starting year of the forecast are also over-written for these year classes by reducing the GMrecruitment by the appropriate number of years of M (as there is no catch for the first few years of age).

WGCSE (2013) reviewed some information on environmental influences on sea bass recruitment which supported a recent reduction in recruitment. Survival of 0-gp and 1-gp sea bass in nursery areas in estuaries and salt-marshes is thought to be enhanced by warmer conditions promoting survival through the first two winters, and increasing the growth rates (Pawson, 1992). WGCSE 2014 presented an argument for choosing a particular recruitment value for the 2012 year class for inclusion in forecasts, based on a consideration of past recruitment in relation to temperature. The data and arguments in WGCSE 2014 should be consulted for an explanation of the logic used.

The format for reporting the recruitment values for the short-term forecast are summarised using an example in Table D1.1., and an example of a short-term forecast input file is given in Table D.1.2.

Table D.1.1. Example of recruitment estimates included in a short-term forecast for sea bass, from IBPBass 2013.

| Year class | SS3 (AGE 0) | GM 2008-2011 | GM 2008-2011 |
| :--- | :--- | :--- | :--- |
| 2011 | 2648 thousand |  |  |
| 2012 |  | 1815 thousand |  |
| 2013 |  | 6057 thousand |  |
| 2014 |  | 6057 thousand |  |

Example input for the short-term catch predictions is in Table D 1.2. The derivation of the inputs is described in Table D 1.3.

Table D.1.2. Example inputs for sea bass short-term forecast, from WGCSE 2014. Inputs for shortterm forecast. $F(5-11)$ is mean for years 2011-2013. Numbers-at-ages $0-2$ in 2014 are adjusted by replacing Stock Synthesis average values in 2012-2014 (years with no recruit deviations estimated) with the short-term (2008-2011) GM in 2012 and the long-term GM in 2013 and 2014, with adjustments for natural mortality. Rules are below table.


Table D 1.3. Derivation of short-term forecast inputs (based on example from IBPBass 2014).

| Input data | Derivation |
| :---: | :---: |
| Starting numbers-at-age 0-16+ in first year (intermediate year) | SS3 output. (N age zero overwritten where necessary by long-term GM, short-term GM or other predicition, reduced by $\mathrm{M}=0.15$ the the require number of years (if no commercial catches) or multiply N at-age in starting year from the assessment by ratio of the replaced recruit value with the SS3 estimate. |
| Recruitment 2014 onwards | Long-term GM, short-termGM or other predictor. |
| Mean wt-at-age in stock | SS3 output |
| Proportion mature (female) | SS3 output |
| Commercial fishery (H-cons) mean F at-age | Average last three years: SS3 output Zs minus $\mathrm{M}=0.15$ and recreational F at-age |
| Commercial fishery (H-cons) mean weight-atage | SS3 output figures on mean yeight in UK, French and other fleets, weighted by SS3 model estimates of landings numbers-at-age for the fleets |
| Recreational removals F at-age | Input values to SS3 (year-invariant), based on commercial lines selectivity) |
| Recreational removals weights-at-age | Output values for UK commercial fleets from final SS3 Run |
| M | 0.15 at all ages |

An example detailed forecast is given in Table D1.4 for the status quo F option, which is the most likely forecast given the absence of any restrictive management controls on effort or landings of sea bass. See WGCSE 2014 for examples of management options tables.

Future forecast routines may be configured in Stock Syntheses, allowing MCMC estimation of confidence limits.

Table D 1.4. Example of detailed short-term forecast (WGCSE 2014).


## E. Biological reference points

## E. 1 Background

The Stock synthesis model currently used fixes stock-recruit steepness at 0.999 . There are insufficient observations at low SSB to suggest the possible steepness of the relationship. This means that MSY reference points cannot be obtained from a plausible stock-recruit relationship and have to be derived from yield-per-recruit. WGCSE 2013 and 2014, and IBPBass 2014 computed yield-per-recruit based biological reference points $\mathrm{F}_{0.1}$ and $\mathrm{F}_{35 \% \mathrm{spr}}$ based on the inputs and outputs of the stock synthesis update assessment. In 2013, with an input M of 0.20 , the $\mathrm{F}(5-11)$ value for $\mathrm{F}_{35} \%$ spr was 0.17 and the $\mathrm{F}_{0.1}$ was 0.18 . In 2014, WGCSE reduced the M to 0.15 , and this caused a reduction in the $\mathrm{F}_{35 \% \text { spr }}$ to 0.13 for combined commercial and recreational landings, and the $\mathrm{F}_{0} .1$ was 0.12. The F reference points therefore scale directly with M , as expected, with some effect of changes in estimated fishery selection patterns and weights-at-age.
The revised SS3 model proposed by IBPBass and applied at WGCSE 2014 now includes an explicit recognition of the possible recreational fishing mortality, albeit as a fixed vector of F at-age following the same selectivity as the commercial line fishery. This leads to a more complex problem in defining BRPs based on yield-per-recruit because the fishing mortality now has separate commercial and recreational F components which can be manipulated separately or simultaneously in a yield-per-recruit analysis. The recreational $F$ vector is indicative only, in the sense that it is conditioned on a total annual recreational removals estimate of $\sim 1500 t$ for recent years, which can be considered as a "plausible scenario" rather than an explicit estimate of recreational F with recreational fishery survey data and their precision included in the model fitting. Different scenarios for total recreational removals affect how total fishing mortality is split between commercial and recreational F, but the combined F estimates are minimally affected as they are driven by the fishery composition data.

## E. 2 MSY BRPs or proxies for sea bass

BRPs for thís assessment based on advice from ICES WKMSYREF2 (ICES, 2014), and recognising that $\mathrm{F}_{\mathrm{MSY}}$ cannot be obtained from a fitted stock-recruit curve include:

1 ) Setting an $\mathrm{F}_{\text {ms }}$ proxy as $\mathrm{F}_{0.1}$ or $\mathrm{F}_{x x \%}$ spr based either on the commercial fishery only, or the combined commercial and recreational F.
Setting a BMSY trigger around a low percentile of the expected range of SSB when fishing at $\mathrm{F}_{\mathrm{mSY}}$.

Unfortunately it has not been possible yet to carry out a full MCMC bootstrap of the sea bass assessment and to propagate this into a forecast period to evaluate the percentiles of expected SSB whilst fishing at FMSY. A concern with sea bass is that recruitment has shown longer term changes in mean recruitment (and hence stock productivity) that appear related to changes in sea temperature at decadal scales. For a biomass reference point, WGCSE 2014 selected the lowest observed SSB (Bloss) of 5250 t as a value for Blim, the limit reference point for SSB. There is therefore no MSY or precautionary reference point for biomass (Table E.2.1).

WGCSE chose $\mathrm{F}_{35 \%}$ spr as a proxy for $\mathrm{F}_{\text {MSY }}$ (value: $\mathrm{F}(5-11)=0.13$ for combined commercial and recreational fishery). As for many stocks, this value is close to $\mathrm{F}_{0.1}$ ( $\mathrm{F}=0.12$ ). (Table E.2.1).

The yield per recruit curve is flat-topped and FmSY in sea bass is poorly defined (Figure E.2.1). Fishing at $\mathrm{F}_{\text {max }}$ implies a yield-per-recruit only marginally higher than at $\mathrm{F}_{35 \% \text { SPR }}$ but requires much higher F , with much larger fishing costs where these are proportional to $F$.

## E. 3 Inclusion of BRP in short-term forecast

For the estimation of Yield-per-recruit and Spawner biomass per recruit, WGCSE 2014 varied the multipliers on both fisheries by the same amount. The resulting yield-perrecruit curve therefore reflects total F and total yield, but does not assume a particular allocation between any fisheries.
For short-term forecasts, WGCSE agreed that the multiplier on the recreational F vector should be maintained at 1.0 for all management options except zero $F$, and only the commercial fishery vector is altered in management options. The relative contribution of commercial and recreational F to the total F will not be the same as the contribution used in calculating the Fmsy based on the mean F for 2011 to 2013. However, the selectivity and weights-at-age for the two fisheries are sufficiently close that the agreed Fmsy can be applied for different combinations of commercial and recreational F with relatively minor differences. This allows managers to select an F relative to Fmsy for a forecast year but decide how the resultant catch forecast can be allocated between the different fishery sectors. This is no different to management of different commercial fishery fleets.

Table E.2.1. BRPs proposed by WGCSE 2014 for sea bass in Areas IV and VII.



Figure E.2.1. Yield and biomass per recruit analysis from WGCSE 2014. conditional on mean pattern of F-at-age for 2011-2013 for the commercial fishery and the vector of F-at-age for recreational fishing, with the same F-multiplier applied to both. The partial YPR by fleet is shown conditional on the relative contribution of the F-at-age in the two fisheries to the combined $F$ being the same as estimated for recent years in the assessment. The $35 \%$ spr is indicated (red line) to show where this occurs on the SPR curve.

## E. 4 Yield-per-recruit reference point calculations

Example inputs for a yield-per-recruit analysis are given in Table E.4.1.1. They are identical to the short-term forecast inputs except that the values are extended to 30 years of age. SS3 provides population estimates out to this age. It is assumed that no fish survive after the 30 thyear, as no fish older than 28 years have been observed historically. The input data are entered into the ICES Standard Plots system on SharePoint, which produces the YPR and SPR plots together with estimates of $\mathrm{F}_{\text {MAX }}$, $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\text {med. }} \mathrm{F}_{35 \% \text { SPR }}$ is not output but can easily be computed using a spreadsheet, checking that the $\mathrm{F}_{0.1}$ and


Table E.4.1.1. Example inputs to yield-per-recruit analysis. $\mathrm{F}(5-11)$ and weights-at-age are the recent three year average. The values for a $16+\mathrm{gp}$ are indicated although using this prevents the mean stock weight to increase in the plus group as F is reduced, leading to a small bias in SSB per recruit at low or zero $F$. Bass in the wild have been recorded up to 28 years old.

| Age | M | Pmat | stock <br> wt (kg) | F(5-11): commercial fleets | F(5-11): recreational fishery | Catch wt (kg): commercial fleets | Catch wt (kg): recreational fleets |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.15 | 0.000 | 0.002 | 0.000 | 0.000 | 0.007 | 0.007 |
| 1 | 0.15 | 0.000 | 0.022 | 0.000 | 0.000 | 0.076 | 0.052 |
| 2 | 0.15 | 0.000 | 0.101 | 0.000 | 0.000 | 0.388 | 0.154 |
| 3 | 0.15 | 0.000 | 0.218 | 0.014 | 0.002 | 0.584 | 0.295 |
| 4 | 0.15 | 0.186 | 0.382 | 0.070 | 0.013 | 0.715 | 0.480 |
| 5 | 0.15 | 0.419 | 0.587 | 0.161 | 0.054 | 0.877 | 0.702 |
| 6 | 0.15 | 0.638 | 0.825 | 0.220 | 0.091 | 1.093 |  |
| 7 | 0.15 | 0.792 | 1.088 | 0.247 | 0.099 | 1.331 | 228 |
| 8 | 0.15 | 0.885 | 1.370 | 0.256 | 0.100 | 1.588 |  |
| 9 | 0.15 | 0.937 | 1.664 | 0.266 | 0.100 | 1.860 | 1.815 |
| 10 | 0.15 | 0.965 | 1.964 | 0.271 | 0.100 | 43 | . 116 |
| 11 | 0.15 | 0.980 | 2.264 | 0.274 | 0.100 | 2.431 | . 416 |
| 12 | 0.15 | 0.989 | 2.562 | 0.275 | 0.100 | . | 2.711 |
| 13 | 0.15 | 0.993 | 2.853 | 0.276 | 0.100 | 3.002 | 2.99 |
| 14 | 0.15 | 0.996 | 3.135 | 0.276 | 0.100 | 3.276 | 75 |
| 15 | 0.15 | 0.998 | 3.406 | 0.276 | 0.100 | 3.540 | 3.540 |
| 16 | 0.15 | 0.998 | 3.665 | 0.276 | 0.100 | 3.791 | 3.793 |
| 17 | 0.15 | 0.999 | 3.911 | 0.277 | . 100 | 4.030 | 4.032 |
| 18 | 0.15 | 0.999 | 4.143 | 0.277 | 0.100 | . 255 | 4.258 |
| 19 | 0.15 | 1.000 | 4.362 |  | 0.100 | 4.466 | 4.469 |
| 20 | 0.15 | 1.000 | 4.566 | 7 | 0.10 | 4.664 | 4.667 |
| 21 | 0.15 | 1.000 | 4.757 | 0.277 | 0.100 | 4.848 | 4.852 |
| 22 | 0.15 | 1.000 | 4.935 | 0.27 | 0.100 | 5.019 | 5.023 |
| 23 | 0.15 | 0 | 5.099 | 0.277 | 0.100 | 5.177 | 5.181 |
| 24 | 0.15 | 1.000 | 5.252 | 0.277 | 0.100 | 5.324 | 5.328 |
| 25 |  | 000 | 393 | 0.277 | 0.100 | 5.459 | 5.464 |
| 26 |  | 1.000 | 5.52 | . 277 | 0.100 | 5.584 | 5.588 |
| 27 | 0.15 | 1.000 | 5.642 | 0.277 | 0.100 | 5.698 | 5.703 |
|  | . 15 | 1.000 | 5.752 | 0.277 | 0.100 | 5.803 | 5.808 |
|  |  | 1.000 | 5.853 | 0.277 | 0.100 | 5.900 | 5.905 |
|  | 0.15 | 1.000 | 6.064 | 0.277 | 0.100 | 6.102 | 6.107 |
| $16+$ | 0.15 | 1.000 | 4.017 | 0.277 | 0.100 | 4.069 | 4.134 |

There is currently no TAC for sea bass, and control of fishing mortality would have to be through other approaches to managing effort on sea bass, and including technical measures to alter selectivity and/or restrict fishing seasonally or spatially. Note that the tclusion of discards in the assessment would alter the reference points and historical series. In the absence of discards, it is difficult to infer benefits to YPR or SSB/R in improving the selectivity patterns of the fleets. There is currently no time-series of recreational fishing catches to monitor the impacts of any management measures, and the frequency and extent of future surveys remains uncertain.

## F. Other Issues

## F.1. Historical overview of previous assessment methods

Previous assessments of sea bass in the IV \& VII area are summarised below.
2007: Pawson et al. 2007. ADMB separable model on UK data; updated 2008 at WGNEW (Kupschus et al., 2008).

2012: IBP-NEW (ICES, 2012). Development of age and length based Stock Synthesis assessment.

2013: WGCSE. Update assessment using IBP-NEW SS model. Recommended inter-benchmark to improve model.

2014: IBP-BASS. Added new CGFS surveys series; removed poorly performing surveys; improved fleet structure and selectivity model; incorporated recreational fishery information; developed forecast and BRPs.
2014: WGCSE. Update assessment using IBP-Bass model.

## E. References

Armstrong, M.J. 2012. Life-history estimates of natural mortality of sea bass around the UK. Working Document: ICES IBP-NEW 2012; October 2012. 3pp.
Armstrong, M.J. and Walmsley, S. 2012a. An evaluation of the bass fleet census and logbook system for estimating annual landings by gear for fishing vessels in England and Wales. Working Document: ICES IBP-NEW 2012; October 2012. 1
Armstrong and Walmsley. 2012b. Age and growth of sea bass sampled around the UK. Working Document: ICES IBP-NEW 2012; October 2012. 15 pp .
Armstrong and Walmsley. 2012c. Maturity of sea bass sampled around the UK. Working Document: ICES IBP-NEW 2012; October 2012. 14pp.
Armstrong, M.J. and Maxwell, D. 2012. Commercial fleet lpuetrends for sea bass around the UK. Working Document: ICES IBP-NEW 2012; October 2012. 29 pp.
Armstrong, M. and Readdy, L. 2014. Effect on sea bass Stock Synthesis model of expanding the UK fishery age compositions to a larger plus group. Working Document, ICES IBPBASS, 2014.

Armstrong, M.J. and Drogou, M. 2014. Seabass fisheries in Europe and their management. Commission request for services SI2. 680348. 83pp.; presented also to STECF July 2014 plenary.
Armstrong M, Brown A, Hargreaves J, Hyder K, Pilgrim-Morrison S, Munday M, Proctor S, Roberts A, Roche N, Williamson K (2013). Sea Angling 2012 - a survey of recreational sea angling activity and economic value in England. Defra report, © Crown copyright 2013.
http://webarchive.nation-
alarchives.gov.uk/20140108121958/http://www.marinemanagement.org.uk/seaangling/index.htm
pin, F., Le Roy, D., Schlaich, Y. 2002. Manuel des protocoles de campagne halieutique: Campagnes CGFS, Système d'information halieutiques - Campagnes à la mer. Ifremer, 09/2001-DRV/RH/DT/AN-NUMERO, 29pp. (in French).
Diodati, P. and R.A. Richards. 1996. Mortality of striped bass hooked and released in salt water. Transactions of the American Fisheries Society. 125: 300-307.

Drogou M et al. 2011. Synthèse des informations disponibles sur le Bar : flottilles, captures, marché. Reflexions autour de mesures de gestion.
Dunn, M.R. and Potten, S. 1994. National Survey of Bass Angling: Report to the Ministry of Agriculture, Fisheries and Food. University of Portsmouth, Centre for the Economics and Management of Aquatic Resources. 45pp + appendices.

Dunn, M., Potten, S., Radford, A. and Whitmarsh, D. 1989. An Economic Appraisal of the Fishery for Bass in England and Wales. Report to the Ministry of Agriculture, Fisheries and Food. University of Portsmouth. 217 pp.

Herfaut J., Levrel H., Drogou M. and Véron G. 2010. Monitoring of recreational fishing of sea bass (Dicentrarchus labrax) in France: output from a dual methodology (telephone survey and diary) ICES CM 2010/R: 05.

ICES. 2001. Report on the ICES Study Group on Bass. CM 2001/ACFM:25, 18 pp.
ICES. 2002. Report on the ICES Study Group on Bass. CM 2002/ACFM:11 ref.G, 59 pp.
ICES. 2004a. Report of the Study Group on Bass, Lowestoft, England, August 2003. ICES Document, CM 2004/ACFM: 04.73 pp .
ICES. 2004b. Report of the Study Group on Bass, By Correspondence. ICES Document, CM 2004/ACFM: 31 Ref G. 56pp.
ICES. 2008. Report of the Working Group on the Assessment of New MoU Species (WGNEW). By Correspondence, ICES CM 2008/ACOM:25. 77 pp.
ICES. 2009. Report of the ICES Workshop on Sampling Methods for Recreational Fisheries (WKSMRF). ICES CM 2009/ACOM: 41.

ICES. 2010 Report of the Planning Group on Recreational Fisheries (PGRFS), 7-11 June 2010, Bergen, Norway. ICES CM 2010/ACOM:34. 168 pp.

ICES. 2011. Report of the Planning Group on Recreational Fisheries (PGRFS). ICES CM 2011/ACOM:23.

ICES. 2012a. Report of the Inter-Benchmark Protocol on New Species (Turbot and Sea bass; IBPNew 2012). ICES CM. 2012/ACOM:45.

ICES. 2012b. Report of the Working Group on Assessment of New MoU Species (WGNEW), 59 March 2012. ICES CM 2012/ACOM:20. 258 pp.
ICES. 2012c. Report of the Working Group on Recreational Fisheries Surveys (WGRFS). ICES CM 2012/ACOM:23. 55 pp .
ICES. 2013a. Report of the Working Group for Celtic Seas Ecoregion (WGCSE), 8-17May 2013, Copenhagen, Denmark. ICES CM 2013/ACOM:12. 1986 pp.
ICES, 2013. Report of the Working Group on Recreational Fisheries Surveys (WGRFS). ICES CM 2013/ACOM:23.
Jennings, S., and Pawson, M. G. 1992. The origin and recruitment of bass, Dicentrarchus labrax, larvae tonursery areas. Journal of the Marine Biological Association of the United Kingdom, 72: 199-212

Kennedy, M. and Fitzmaurice, P. 1968. Occurrence of eggs of bass, Dicentrarchus labrax, on the southern coasts of Ireland. Journal of the Marine Biological Association of the U.K., 48: 585592.

Kennedy, M. and Fitzmaurice, P. 1972. The biology of the bass, Dicentrarchus labrax in Irish waters. Journal of the Marine Biological Association of the United Kingdom 52: 557-597.

Kupschus, S., Smith, M. T., Walmsley, S. A. 2008. Annex 2: Working Document. An update of the UK bass assessments 2007. Report of the Working Group on the Assessment of New MoU Species (WGNEW). By Correspondence, ICES CM 2008/ACOM:25. 77 pp.
Lam Hoai Thong. 1970. Contribution à l'étude des Bars de la région des Sables d'Olonne. Trav. Fac. Sci. Rennes, Ser. Océanogr. Biol., 3: 39-68.

Lancaster, J. E. 1991. The feeding ecology of juvenile bass Dicentrarchus labrax (L.). PhD thesis, University College of Swansea, 281 pp.
Laurec et al. 2012. Analysis of length distribution in sea bass for a given read age. WD to IBPNEW 2012.

Mahé, K., Holmes, A., Huet, J., Sévin, K., Elleboode, R. 2012. Report of the Seabass (Dicentrachus labrax) Otolith and Scale Exchange Scheme 2011, 16 pp.

Masski, H. 1998. Identification de Frayères et Etude des Structures de Population de Turbot (Psetta maxima L.) et du Bar (Dicentrarchus labrax L.) en Manche Ouest et dans les Zones Avoisinantes. Thèse présentée a la Faculte des Sciences de Brest. Universite de Bretagne Occdentale. 136pp + annexes.

Mayer, I., Shackley, S.E. and Witthames, P.R. 1990. Aspects of the reproductive biology of the bass, Dicentrarchus labrax L. II. Fecundity and pattern of oocyte development. J. Fish Biol. 36:141-148.

Methot, R.D. 2000. Technical Description of the Stock Synthesis Assessment Program. National Marine Fisheries Service, Seattle, WA. NOAA Tech Memo. NMFS-NWFSC-43: 46 pp.

Methot, R.D. 2011. User Manual for Stock Synthesis, Model Version 3.23b. NOAA Fisheries Service, Seattle. 167 pp.

Pawson, M. G. 1992. Climatic influences on the spawning success, growth and recruitment of bass (Dicentrarchus labrax L.) in British Waters. ICES mar. Science Symp. 195: 388-392.
Pawson, M. G., Kupschus, S. and Pickett, G. D. 2007a. The status of sea bass (Dicentrarchus labrax) stocks around England and Wales, derived using a separable catch-at-age model, and implications for fisheries management. ICES Journal of Marine Science 64, 346-356.

Pawson, M. G., Kelley, D. F. and Pickett, G. D. 1987. The distribution and migrations of bass Dicentrarchus labrax L. in waters around England and Wales as shown by tagging. J. mar. biol. Ass. UK, 67: 183-217.

Pawson, M. G., and Pickett, G. D. 1996. The annual pattern of condition and maturity in bass (Dicentrarchus labrax L) in waters around the UK. Journal of the Marine Biological Association of the United Kingdom, 76: 107-126.

Pickett, G.D. 1990. Assessment of the UK bass fishery using a log-book-based catch recording system. Fish. Res. Tech. Rep., MAFF Direct. Fish Res., Lowestoft 90: 30pp.
Pickett, G. D., and Pawson, M. G. 1994. Bass. Biology, Exploitation and Management. Chapman \& Hall, London, Fish and Fisheries Series, 12. 358 pp.

Quirijns, F. and Bierman, S. 2012. Growth and maturity of sea bass sampled around the Netherlands. Working Document: ICES IBP-NEW 2012; October 2012. 9pp.
Reynolds, W. J., Lancaster, J. E. and Pawson, M. G. 2003. Patterns of spawning and recruitment of bass to Bristol Channel nurseries in relation to the 1996 "Sea Empress" oil spill. J. Mar. Biol. Assoc. UK, 83: 1163-1170.

Rocklin D, Levre1 H, Drogou M, Herfaut J, Veron G. 2014. Combining Telephone Surveys and Fishing Catches Self-Report: The French Sea Bass Recreational Fishery Assessment. PLoS ONE 9(1): e87271. doi:10.1371/journal.pone.0087271.
uert, B., 1972. Contribution à l'étude du bar Dicentrarchus labrax (L.) des reservoirs à poissons de la région d'Arcachon. Th. 3ème year: Faculté des Sciences.

Thompson, B. M., Harrop, R. T. 1987. The distribution and abundance of bass (Dicentrarchus labrax) eggs and larvae in the English Channel and Southern North Sea. Journal of the Marine Biological Association of the United Kingdom, 67, 263-274.

Tulp, I., Bolle, L. J. and Rijnsdorp, A.D. 2008. Signals from the shallows: In search of common patterns in long-term trends in Dutch estuarine and coastal fish. Journal of Sea Research 60 (1-2), pp. 54-73.
van der Hammen, T and de Graaf, M. 2012. Recreational fishery in the Netherlands: catch estimates of cod (Gadus morhua) and eel (Anguilla anguilla) in 2010. IMARES Wageningen UR, Report Number C014/12, 61 pp.
van der Hammen T, and de Graaf, M. 2015. Recreational fisheries in the Netherlands: analyses of the 2012-2013 logbook survey, 2013 online screening survey and 2013 random digit dialing survey. Imares C042/15, pp. 55.
van der Hammen, T., Poos, J. J., van Overzee H. M.J., Heessen, H. J.L. and Rijnsdorp A. D. 2013. Data evaluation of data limited stocks: Horse mackerel, Sea bass, Greater Silver Smelt, Turbot and Brill. Imares report number C166/13.
van Beek, F. A., Rijnsdorp, A.D., de Clerck, R. 1989. Monitoring juvenile stocks of flatfish in the Wadden Sea and the coastal areas of the southeastern North Sea. Helgoländer Meeresuntersuchungen, 43 (3-4), pp. 461-477. doi: 10.1007/BF02365904.

Walmsley, S. and Armstrong, M. 2012. The UK commercial bass fishery in 2010. Working Document to ICES WGNEW 2012. August 2011.

## Appendix 1

Content of Stock Synthesis Control File (BassIVVII.ctl) used at WGCSE 2014. The file was originally built from one used for other species, and some comments remain that are for those assessments. Rows preceded by \# are skipped, and are greyed out.
\#C growth parameters are estimated
\#_SS-V3.04-safe;_09/09/09;_Stock_Synthesis_by_Richard_Methot_(NOAA);_using_ADMB_7.0.1
1 \#_N_Growth_Patterns
1 \#_N_Morphs_Within_GrowthPattern(GP)
\#_Cond 1
\#_Morph_between/within_stdev_ratio (no read if N
\#_Cond 1
\#vector_Morphdist_(-1_in_first_val_gives
\#
\#\#1 \# N recruitment designs goes here if N_GP*nseas*area $>1$ \#here 1 sp, 4 seasons, 1 area
\#\#0 \# placeholder for recruitment interaction reque
\#GP seas area for each recruitment assignment
\#\#111 \# example recruitment design element for

\#_Cond 1.0 \# first age that moves (real age a begin of season, not integer) also cond on do_migration>0
\#_Cond 1112410 \# example move defintion for seas=1, morph=1, source=1 dest=2, age $1=4$, age2=10
\#

\# begin and end years olocks in first pattern
0.5\#_fracfemale \#? Note sex ratio in bass increases with length.

3 \#_natM_type:_0=1Parm; 1=N_breakpoints;_2=Lorenzen;_3=agespecific;_4=agespec_withseasinterpolate

| 0.150 | 0.150 | 0.150 | 0.152 | 0.163 | 0.204 | 0.241 | 0.249 | 0.250 | 0.250 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 |
|  | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 |


| 0.250 | 0.250 | 0.250 |
| :--- | :--- | :--- |

\#_no additional input for selected M option; read 1P per morph
1 \# GrowthModel: 1=vonBert with L1\&L2; 2=Richards with L1\&L2; 3=not implemented; 4=not implemented \#note - maguire et al 2008 pg 1270, Downloaded from icesjms.oxfordjournals.org at ICES on October 17, 2011

2 \#_Growth_Age_for_L1
999 \#_Growth_Age_for_L2 (999 to use as Linf)
0 \#_SD_add_to_LAA (set to 0.1 for SS2 V1.x compatibility)

0 \#_CV_Growth_Pattern: $0 \mathrm{CV}=\mathrm{f}(\mathrm{LAA}) ; 1 \mathrm{CV}=\mathrm{F}(\mathrm{A}) ; 2 \mathrm{SD}=\mathrm{F}(\mathrm{LAA}) ; 3 \mathrm{SD}=\mathrm{F}(\mathrm{A})$
1 \#_maturity_option: 1=length logistic; 2=age logistic; 3=read age-maturity matrix by growth_pattern; $4=$ read age-fecundity; $5=$ read fec and wt from wtatage.ss
\#_placeholder for empirical age-maturity by growth pattern
4 \#_First_Mature_Age
1 \#_fecundity option:(1)eggs=Wt*(a+b*Wt);(2)eggs=a*L^b;(3)eggs=a*Wt^b
0 \#_hermaphroditism option: $0=$ none; $1=$ age-specific fxn
1 \#_parameter_offset_approach (1=none, 2= M, G, CV_G as offset from female-GP1, 3=like SS2 V1.x)
1 \#_env/block/dev_adjust_method (1=standard; 2=logistic transform keeps in base parm bounds; $3=$ standard w/ no bound check)
\#
\#_growth_parms
\#_LO HI INIT PRIOR PR_type SD PHASE env-var use_dev dev miny dev_matr dev_stddev

0.01 0.2 0.09699 0.09699-1 0.05-3 0000000 \# VonBert K_GP_1
0.005 0.5 0.2 0.2-1 0.8-60000000 \# CV_young_GP_1
0.005 0.5 0.1 0.1-1 0.8-6 0000000 \# CV_old_GP_1
-1 $10.000012960 .00001296-10.05-30000000$ \# Wtlen_1
24 2.969 2.969-1 0.05-3 0000000 \# Wtlen_2
305040.649 40.649-1 5-3000 0000 \# Mat50\%
-5 1-0.33349-0.33349-1 0.03764-30000000 \# Mat_slope
-3311-1 0.8-30000000\#Eg/gm_inter
-3300-1 0.8-30000000 \# Eg/gm_slope_wt
0000-1 0-30000000 \# RecrDist_GP_1
0000-10-30000000 \# RecrDist_Area_1
0000-10-40000000 \# RecrDist_Seas_1
0000-10-40000000 \# CohortGrowDev
\#
\#_Cond 0 \#custom_MG-env_setup (0/1)
\#_Cond -2 2 0 0-1 99-2 \#_placeholder when no MG-environ parameters
\#
\#_Cond 0 \#custom_MG-block_setup (0/1)
\#_Cond -2 2 0 0-1 99-2 \#_placeholder when no MG-block parameters
\#_Cond No MG parm trends
\#
\#_seasonal_effects_on_biology_parms



## 0000 \# Survey AutBass 4 <br> 0000 \# Survey CGFS1

\#_Cond 0 \#_If q has random component, then $0=$ read one parm for each fleet with random q; $1=$ read a parm for each year of index
\#_Q_parms(if_any)
\# LO HI INIT PRIOR PR_type SD PHASE
\#-15 500-11-1 \# Q_base_2_SURVEY EVHOE
\#
\#_size_selex_types
0000 \# UKTrawl_Nets \#_RDM now all fleets have size selectivit
0000 \# UKMidwater
1000 \# French
15003 \# Other
0000 \# AutBass2
0000 \# AutBass3
0000 \# AutBass4
24000 \# CGFS1
\#
\#_age_selex_types
\#_Pattern $\qquad$ Male Special

20000 \# 1 UKTrawl
$12000 \#$ 2 UKMidwater
10000 \# 3 French
15003 \# 4 Other
11000 \# 5 AutBass
11000 \# 6-AutBass 3
11000 \# 7 AutBass4
10000 \# 8 CGFS1
\#_LO HI INIT PRIOR PR_type SD PHASE env-var use_dev dev_minyr dev_maxyr dev_stddev Block Block_Fxn

2091 38.8836 30-1 0.120000000 \# SizeSel_5P_1_French
0.01305 .185145 -1 0.0130000000 \# SizeSel_5P_2_French

### 1.0 5.0 3.3 3.3-1 0.053000000 \# SizeSel_2P_3_CGFS1 <br> 3.0 6.0 4.4 4.4-1 0.053000000 \# SizeSel_2P_4_CGFS1 <br> -8.0 9.0-8 -8 -1 0.0520000000 \# SizeSel_2P_5_CGFS1 <br> -5.0 9.0-1-1-1 0.0520000000 \# SizeSel_2P_6_CGFS1

0 15.8 7.4 7.4-1 0.05 20000000 \# PEAK value SizeSel_1P_1_OTB -6.0 4.0-6.0-6.0-1 0.053000000 \# TOP logistic SizeSel_1P_1_OTB -3 9.0 0.3 0.3-1 0.0530000000 \# WIDTH exp SizeSel_1P_1_OTB -1.0 9.0-0.8-0.8-1 0.0530000000 \# WIDTH exp SizeSel_1P_1_OTB -10.0 9.0-5 -5 -1 0.0520000000 \# INIT logistic SizeSel_1P_1_OTB -5.0 9.01.8 1.8-1 0.0520000000 \# FINAL logistic SizeSel_1P_1_OTB
$0167.07 .0-10.0520000000$ \# SizeSel_2P_1_MWT $0.01307 .023995-10.0530000000$ \# SizeSel_2P_

2222-199-30000000 \# AgeSel_10P_1_Autumn2 min age 2222-199-30000000 \# AgeSel_10P_2_Autumn 2 maxage

3333-199-30000000 \# AgeSel_11P_1_Autumn 3 min age 3333-199-30000000 \#AgeSel_11P_2_Autumn 3 max age

4444-199-30000000 \# AgeSel_12P_1_Autumn 4 min age 4444-199-30000000\# AgeSel_12P_2_Autumn 4 max age

__placeholder when no enviro fxns
x parm trends
Cond -4 \# placeholder for selparm_Dev_Phase
/block/dev_adjust_method (1=standard; 2=logistic trans to keep in base parm bounds; 3=standard w/ no bound check)
\#
\# Tag loss and Tag reporting parameters go next
0 \# TG_custom: $0=$ no read; $1=$ read if tags exist
\#_Cond -661120.01-40000000 \#_placeholder if no parameters \#

1 \#_Variance_adjustments_to_input_values
\#_fleet/svy: 12345678
00000000 \#_add_to_survey_CV

00000000 \#_add_to_discard_stddev
00000000 \#_add_to_bodywt_CV
11111111 \#_mult_by_lencomp_N
11111111 \#_mult_by_agecomp_N
11111111 \#_mult_by_size-at-age_N
\#
3 \#_maxlambdaphase
1 \#_sd_offset
\#
0 \#8 \# number of changes to make to default Lambdas (default value is 1.0)
\# Like_comp codes: $1=$ surv; $2=$ disc; $3=$ mnwt; $4=$ length; $5=$ age; $6=$ SizeFreq; $7=$ sizeage; $8=$ catch;
\# 9=init_equ_catch; 10=recrdev; 11=parm_prior; 12=parm_dev; 13=Crashßen; 14=Morphcomp; 15=Tag-comp; 16=Tag-negbin
\#like_comp fleet/survey phase value sizefreq_method
\# 5110.11 \#_RDM reduce emphasis on age comp
\# 5210.11
\#5310.11
\# 5410.11
\# 7110.11
\# 7210.11
\# 7310.11
\# 7410.11
\#
\# 11 1 \#_CPLE/syrvey:_3
<1111 \#_lencomp:_2
\# 1111 \#_agecomp:_1
\# 1111 \#_agecomp:_2
\# 0000 \#_agecomp:_3
\# 1111 \#_size-age:_1
\# 1111 \#_size-age:_2
\# 0000 \#_size-age:_3
\# 1111 \#_init_equ_catch
\# 1111 \#_recruitments
\# 1111 \#_parameter-priors

## \# 1111 \#_parameter-dev-vectors

## \# 1111 \#_crashPenLambda

0 \# ( $0 / 1$ ) read specs for more stddev reporting
\# 1 1-1 5 1 5 1-1 5 \# selex type, len/age, year, N selex bins, Growth pattern, N growth ages, NatAge_area(-1 for all), NatAge_yr, N Natages
\# 515253543 \# vector with selex std bin picks (-1 in first bin to self-generate)
\# 12142640 \# vector with growth std bin picks (-1 in first bin to self-generate)
\# 12142640 \# vector with NatAge std bin picks (-1 in first bin to self-generate)
999

## Appendix 2

Content of Stock Synthesis Data file (BassIVVII.dat) used at WGCSE 2014. The file was originally built from one used for other species, and some comments remain that are for those assessments. Rows preceded by \# are skipped, and are greyed out.
\#C Sea bass IV VII input data file
1985 \# _styr lr
2014 \# _endyr lr
1 \# _nseasons lr number of quarters in a year
12 \# _nmonths per season lr months in each quarter
1 \# _spawn_seas ? which season is the spawning season
4 \# _Nfleet \#lr
4 \# _Nsurveys \#lr
1 \# _N_areas \# lr
UKOTB_Nets_Lines\%UKMWT\%French\%Other\%AutBass2\%AutBass3\%AutBass4\%CGFS1
-1-1-1-1 0.83 0.83 0.83 0.75 \# _surveytiming_in_season
11111111 \# _area_assignments_for_each_fishery_and_survey
1111 \# _units of catch: 1=bio; 2=num
0.10 .10 .10 .1 \# _se of $\log$ (catch) only used for init_eq_catch and for Fmethod 2 and 3

1 \#_Ngenders \# lr (?if split by male female then 2? 1r)
30 \#_Nages \# lr
48.70 .51709 .68 118.76 \#CF _init_equil_catch_for_each_fishery ( $82 \%$ of 1985 value)

29 \# _N_lines_of_catch_to_read \#lr updated WGCSE 2014
\# _catch_biomass(tons):_columns_are_fisheries units, year,season Updated WGCSE 2014
$59.72 \quad 0.62 \quad \begin{array}{llll}870.38 & 145.61 & 1985 & 1\end{array}$
-


| 1180.34 | 17.27 | 1986 | 1 |
| :--- | :--- | :--- | :--- |
| 1839.69 | 20.9 | 1987 | 1 |
| 1027.77 | 38.61 | 1988 | 1 |
| 916.96 | 53.14 | 1989 | 1 |
| 849.03 | 25.2 | 1990 | 1 |
| 970.72 | 17.32 | 1991 | 1 |
| 1000.57 | 37.07 | 1992 | 1 |
| 979.35 | 48.26 | 1993 | 1 |
| 786.31 | 59.74 | 1994 | 1 |
| 1056.62 | 109.63 | 1995 | 1 |
| 2395.39 | 82.31 | 1996 | 1 |
| 1984.07 | 91.37 | 1997 | 1 |
| 1773.01 | 143.47 | 1998 | 1 |





| 1995 | 1 | 8 | 1021740 | 0.3 |
| :--- | :--- | :--- | :--- | :--- |
| 1996 | 1 | 8 | 1224238 | 0.3 |
| 1997 | 1 | 8 | 1817599 | 0.3 |
| 1998 | 1 | 8 | 2531044 | 0.3 |
| 1999 | 1 | 8 | 1642270 | 0.3 |
| 2000 | 1 | 8 | 2570996 | 0.3 |
| 2001 | 1 | 8 | 3150674 | 0.3 |
| 2002 | 1 | 8 | 3872427 | 0.3 |
| 2003 | 1 | 8 | 8739057 | 0.3 |
| 2004 | 1 | 8 | 3598440 | 0.3 |
| 2005 | 1 | 8 | 3005317 | 0.3 |
| 2006 | 1 | 8 | 5517999 | 0.3 |
| 2007 | 1 | 8 | 3661314 | 0.3 |
| 2008 | 1 | 8 | 6468841 | 0.3 |
| 2009 | 1 | 8 | 2564696 | 0.3 |
| 2010 | 1 | 8 | 1804537 | 0.3 |
| 2011 | 1 | 8 | 1513745 | 0.3 |
| 2012 | 1 | 8 | 2034554 | 0.3 |
| 2013 | 1 | 8 | 995987 | 0.3 |
| 0 \# _N_fleets_with_discard |  |  |  |  |
| 0 \# N discard obs |  |  |  |  |
| 0 \# _N_meanbodywt_obs |  |  |  |  |
| 30 | \#_DF_for_meanbodywt_T-distribution_like |  |  |  |
| 3 \# lengthbin method: 1=use databins; $2=$ generate from binwidth,min,max below; 3=read vector |  |  |  |  |
| 49 \# number of population length bins to be read |  |  |  |  |



41 \#_N_LengthBins

| 14 | 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 38 | 40 | 42 | 44 | 46 | 48 | 50 | 52 | 54 | 56 |  |
|  | 58 | 60 | 62 | 64 | 66 | 68 | 70 | 72 | 74 | 76 |  |
|  | 78 | 80 | 82 | 84 | 86 | 88 | 90 | 92 | 94 |  |  |

40 \#_N_Length_obs \#lr Updated WGCSE 2014
\#Yr Season Flt/Svy Gender Part Nsamp datavector(female-male)
\#\#\#\# LANDINGS OF COMMERCIAL FLEETS (1 TO 5), ORDERED BY YEAR AND QUARTER:
\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#


|  | 36547 | 57472 | 24016 | 21415 | 27466 | 20198 | 12083 | 7551 | 979 | 1765 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 264 | 1004 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| 2010 | 1 | 3 | 0 | 2 | 200 | 0 | 0 | 0 | 717 | 0 |
|  | 0 | 0 | 0 | 0 | 9811 | 28290 | 169311 | 177571 | 182105 |  |
|  | 283064 | 251956 | 230227 | 188149 | 186310 | 109212 | 120550 | 71590 | 62211 |  |
|  | 31544 | 19076 | 62005 | 26388 | 9340 | 8541 | 29128 | 1884 | 2114 | 182 |
|  | 5525 | 6097 | 863 | 0 | 1207 | 0 | 0 | 0 |  |  |
| 2011 | 1 | 3 | 0 | 2 | 200 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1976 | 13885 | 57121 | 87842 | 128838 |  |
|  | 187586 | 201447 | 199487 | 194697 | 145447 | 124239 | 92526 | 72471 | 46869 |  |
|  | 31690 | 19998 | 17624 | 14720 | 7906 | 6114 | 2082 | 1163 | 1096 | 476 |
|  | 148 | 104 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| 2012 | 1 | 3 | 0 | 2 | 200 | 0 | 0 |  |  |  |
|  | 0 | 0 | 1219 | 0 | 1583 | 6518 | 85760 | 17251 | 140273 |  |
|  | 147895 | 162333 | 180752 | 158490 | 130759 | 107214 | 90638 | 78934 | 54869 |  |
|  | 35387 | 33085 | 17714 | 15170 | 9374 | 8114 | 4147 | 2313 | 1540 | 1134 |
|  | 282 | 451 | 29 | 27 | 0 | 0 |  |  |  |  |
| 2013 | 1 | 3 | 0 | 2 | 200 |  | 0 |  |  | 0 |
|  | 0 | 0 | 0 | 239 | 0 | 1617 | 26153 | 0687 | 272401 |  |
|  | 311983 | 234498 | 270538 | 187922 | 166217 | 140323 | 115326 | 118018 | 78259 |  |
|  | 132959 | 61538 | 79165 | 46759 | 31584 | 32919 | 21221 | 11918 | 3418 | 2153 |
|  | 236 | 2743 | 0 | 193 | 0 |  |  | 0 |  |  |
| -1988 |  | 1 | 8 | 0 |  |  | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  | 0 | 58885 | 0 |  |
|  | 76121 | 20143 | 0 | 0 |  | 20143 | 0 | 39616 | 0 |  |
|  | 15434 | 15434 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 |  |
| -1989-1990 |  |  |  | 0 |  | 3 | 0 | 0 | 0 | 0 |
|  | 0 |  |  |  | 0 | 22235 | 0 | 22235 | 0 |  |
|  | 11011 |  |  |  | 0 | 22235 | 0 | 0 | 0 | 0 |
|  | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  |  |  | 0 | 0 | 0 | 0 | 0 |  |  |
|  |  |  |  | 0 | 0 | 8 | 0 | 0 | 32733 |  |
|  |  | 120021 | 96615 | 222797 | 345962 | 107244 | 0 | 0 | 18463 | 0 |
|  |  |  | 6106 | 23906 | 0 | 17254 | 17254 | 17254 | 0 | 0 |
|  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  |  | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
|  |  | 8 | 0 | 0 | 19 | 0 | 0 | 0 | 0 | 0 |
|  | 304624 | 1719409 | 1119677 | 531390 | 110365 | 162253 | 126980 | 45607 | 16723 |  |
|  | 16703 | 10625 | 0 | 16885 | 0 | 5352 | 30272 | 16885 | 0 | 0 |
|  | 0 | 0 | 0 | 16885 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| 1992 | 1 | 8 | 0 | 0 | 23 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 63079 | 287513 | 510612 | 638459 | 530944 | 309090 | 78328 |  |
|  | 79289 | 26490 | 42201 | 0 | 0 | 15138 | 5247 | 0 | 10494 | 0 |
|  | 0 | 0 | 15853 | 0 | 0 | 5247 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| 1993 | 1 | 8 | 0 | 0 | 21 | 0 | 0 | 0 | 0 | 0 |
|  | 5551 | 38106 | 121556 | 170213 | 347228 | 237317 | 321637 | 247396 | 266632 |  |
|  | 204261 | 133186 | 158039 | 17036 | 0 | 0 | 14798 | 0 | 0 |  |
|  | 16962 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |



|  | 129278 | 137035 | 115126 | 31042 | 39375 | 29156 | 18125 | 9347 | 0 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 61811 | 2692 | 2692 | 9347 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 9347 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| 2005 | 1 | 8 | 0 | 0 | 40 | 0 | 0 | 0 | 0 |  |
|  | 15565 | 0 | 210994 | 418270 | 315562 | 267295 | 354515 | 215506 | 225666 |  |
|  | 181642 | 106404 | 138927 | 146271 | 109108 | 98370 | 34142 | 9487 | 52772 |  |
|  | 51553 | 21543 | 10751 | 10751 | 0 | 0 | 10223 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 2006 | 1 | 8 | 0 | 0 | 36 | 0 | 0 | 0 | 0 | 0 |
|  | 9814 | 197122 | 866753 | 893641 | 788190 | 1120577 | 531053 | 311464 | 137430 |  |
|  | 190156 | 96298 | 71441 | 84223 | 86273 | 31873 | 22996 | 25308 | 9628 | 9814 |
|  | 3032 | 0 | 9814 | 0 | 0 | 11099 | 0 |  |  |  |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| 2007 | 1 | 8 | 0 | 0 | 33 | 0 |  |  |  |  |
|  | 0 | 47391 | 407308 | 386452 | 394375 | 609358 | 479157 | 314000 | 198 |  |
|  | 169994 | 170481 | 111148 | 80056 | 57131 | 123245 | 5737 | 18311 |  |  |
|  | 11929 | 0 | 10662 | 5331 | 19351 | 0 |  |  |  | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |
| 2008 | 1 | 8 | 0 | 0 | 40 |  |  |  | 0 |  |
|  | 10729 | 52876 | 128066 | 266412 | 229414 | 731914 | 1661470 | 931172 | 880632 |  |
|  | 758996 | 224765 | 208707 | 112521 | 79753 | 13259 | 72807 | 14967 | 28587 |  |
|  | 19628 | 0 | 0 | 0 | 42166 |  |  | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  | 0 | 0 |  |
| 2009 | 1 | 8 | 0 | 0 |  |  | 0 | 0 | 0 | 0 |
|  | 0 | 21413 | 77225 | 253598 | 804 | 279160 | 234313 | 217371 | 238780 |  |
|  | 238652 | 168707 | 110777 | 82775 | 99010 | 56555 | 38480 | 0 | 21413 |  |
|  | 12158 | 0 |  | 0 | 12249 | 0 | 0 | 15256 | 0 | 0 |
|  | 0 | 0 |  | 0 |  | 0 | 0 | 0 |  |  |
| 2010 | 1 |  |  |  | 30 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 661 | 79020 | 106240 | 232673 | 118288 | 117886 | 316567 | 207944 |  |
|  | 97593 | 0810 | 191255 | 62479 | 14580 | 62090 | 0 | 22711 | 0 |  |
|  | 580 |  | 4580 | 0 | 0 | 0 | 0 | 14580 | 0 | 0 |
|  |  |  |  | 0 | 0 | 0 | 0 | 0 |  |  |
|  |  |  | 0 | 0 | 27 | 0 | 0 | 0 | 0 | 0 |
|  |  |  | 0 | 45451 | 95055 | 204438 | 227355 | 208472 | 118040 |  |
|  |  | 68027 | 88143 | 90627 | 60054 | 47411 | 9869 | 0 | 33403 |  |
|  | 36907 | 46184 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
|  | 1 | 8 | 0 | 0 | 25 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 29922 | 15009 | 161369 | 400458 | 247182 | 200887 | 183315 | 84361 |  |
|  | 197280 | 103031 | 86442 | 77382 | 31957 | 43195 | 19956 | 39781 | 11851 |  |
|  | 40610 | 19956 | 20305 | 20305 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| 2013 | 1 | 8 | 0 | 0 | 19 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 42825 | 113195 | 62476 | 113528 | 127750 | 96177 | 135950 | 70592 |  |
|  | 33717 | 44279 | 61410 | 11237 | 17382 | 17382 | 48087 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |

## 012345678910111213141516

1 \#_N_ageerror_definitions

| 0.5 | 1.5 | 2.5 | 3.5 | 4.5 | 5.5 | 6.5 | 7.5 | 8.5 | 9.5 | 10.5 | 11.5 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  | 12.5 | 13.5 | 14.5 | 15.5 | 16.5 | 17.5 | 18.5 | 19.5 | 20.5 | 21.5 |
|  |  | 22.5 | 23.5 | 24.5 | 25.5 | 26.5 | 27.5 | 28.5 | 29.5 | 30.5 |  |
| 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
|  |  | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
|  |  | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |  |

45 \#_N_Agecomp_obs Updated WGCSE 2014
1 \#_Lbin_method: 1=poplenbins; 2=datalenbins; 3=lengths
0 \#_combine males into females at or below this bin number \#Yr Seas Flt/Svy Gender Part Ageerr Lbin_lo Lbin_hi Nsamp datavector(female-male)


| 1996 | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 100 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 145 | 9340 | 41876 | 32357 | 89255 | 251801 | 12051 | 6450 | 352 | 794 |
|  | 970 | 4055 | 2652 | 1238 | 3457 |  |  |  |  |  |
| 1997 | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 100 | 0 | 0 |
|  | 330 | 4027 | 10393 | 73598 | 50958 | 51074 | 186353 | 11996 | 5953 | 927 |
|  | 955 | 589 | 4462 | 3208 | 4262 |  |  |  |  |  |
| 1998 | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 100 | 0 | 0 |
|  | 0 | 13955 | 58518 | 55500 | 105347 | 35312 | 26571 | 71835 | 5360 | 1616 |
|  | 202 | 329 | 1167 | 1931 | 2542 |  |  |  |  |  |
| 1999 | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 100 | 0 | 0 |
|  | 346 | 140 | 97612 | 160695 | 56563 | 50986 | 15824 | 13035 |  | 3277 |
|  | 2153 | 260 | 807 | 548 | 3431 |  |  |  |  |  |
| 2000 | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 1 |  |  |
|  | 0 | 9389 | 2804 | 168794 | 84314 | 17784 | 18662 | 7809 | 8269 |  |
|  | 15820 | 1410 | 379 | 26 | 90 | 1105 |  |  |  |  |
| 2001 | 1 | 1 | 0 | 0 | 1 | -1 |  |  |  | 0 |
|  | 488 | 9410 | 80244 | 32923 | 123218 | 30484 | 1073 | 12972 | 8 |  |
|  | 12042 | 13359 | 1464 | 1157 | 112 | 1196 |  |  |  |  |
| 2002 | 1 | 1 | 0 | 0 | 1 | -1 |  |  |  |  |
|  | 522 | $15068$ | $55187$ | $284390$ | 21080 | 92341 | 24876 | 10798 | 12961 | 4029 |
|  | 5697 | 14855 | 823 | 348 |  |  |  |  |  |  |
| 2003 | 1 | 1 | 0 |  |  |  |  | 100 | 0 | 0 |
|  | 0 | 6054 | 71838 | 68737 | 1819 | 9 | 53668 | 19350 | 6287 | 4653 |
|  | 3870 | 2998 | 6685 | 1071 |  |  |  |  |  |  |
| 2004 | 1 | 1 |  | 0 |  | -1 | -1 | 100 | 0 | 0 |
|  | 0 | 4868 | 44567 | 223462 | 84721 | 117748 | 4530 | 16798 | 8033 | 3022 |
|  | 2016 | 2195 | 728 | 2747 |  |  |  |  |  |  |
| 20052006 | 1 |  |  |  | 1 | -1 | -1 | 100 |  | 0 |
|  | 0 |  | 107126 | 69643 | 117337 | 40220 | 59321 | 6590 | 15664 | 2820 |
|  | 15 | 79 |  |  |  |  |  |  |  |  |
|  |  |  |  | 0 | 1 | -1 | -1 | 100 | 0 | 0 |
|  |  | 8742 | 133803 | 158129 | 61442 | 88113 | 25791 | 41969 | 3889 | 6116 |
|  |  | 138 | 591 | 930 | 3977 |  |  |  |  |  |
|  |  |  | 0 | 0 | 1 | -1 | -1 | 100 |  | 0 |
|  |  |  | 41363 | 208202 | 127765 | 59388 | 51729 | 20972 | 23751 |  |
|  | 15466 | 7284 | 5952 | 87 | 535 | 849 |  |  |  |  |
|  | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 100 |  | 0 |
|  | 0 | 10123 | 119526 | 326706 | 183313 | 69770 | 31832 | 29029 | 13352 |  |
|  | 10641 | 2043 | 1730 | 4019 | 478 | 1703 |  |  |  |  |
| 2009 | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 100 | 0 | 0 |
|  | 0 | 2816 | 78121 | 183363 | 197627 | 86301 | 29037 | 22777 | 12978 |  |
|  | 12790 | 7765 | 2962 | 2268 | 0 | 225 |  |  |  |  |
| 2010 | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 100 | 0 |  |
|  | 0 | 1067 | 77792 | 188027 | 162311 | 111066 | 42029 | 15853 | 8777 | 8349 |
|  | 3153 | 2497 | 2515 | 577 | 1761 |  |  |  |  |  |
| 2011 | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 100 | 0 | 0 |
|  | 0 | 841 | 32410 | 131931 | 111148 | 84747 | 67313 | 41381 | 15244 |  |
|  | 16038 | 8388 | 4829 | 4086 | 1212 | 2092 |  |  |  |  |


| 2012 | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 100 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1690 | 15902 | 187155 | 248719 | 82602 | 59754 | 47075 | 24028 |  |
|  | 16085 | 6655 | 6120 | 4371 | 1056 | 1634 |  |  |  |  |
| 2013 | 1 | 1 | 0 | 0 | 1 | -1 | -1 | 100 | 0 | 0 |
|  | 0 | 0 | 46413 | 67860 | 219457 | 150956 | 51911 | 45186 | 33248 |  |
|  | 28066 | 17913 | 6654 | 5558 | 4380 | 4008 |  |  |  |  |
| 1996 | 1 | 2 | 0 | 0 | 1 | -1 | -1 | 50 | 0 | 0 |
|  | 0 | 0 | 289 | 796 | 3892 | 71665 | 5583 | 1648 | 21 | 334 |
|  | 154 | 622 | 485 | 199 | 560 |  |  |  |  |  |
| 1998 | 1 | 2 | 0 | 0 | 1 | -1 | -1 | 50 | 0 | 0370 |
|  | 0 | 0 | 264 | 6405 | 12691 | 9161 | 8714 | 26925 |  |  |
|  | 100 | 57 | 128 | 957 | 614 |  |  |  |  |  |
| 1999 | 1 | 2 | 0 | 0 | 1 | -1 | -1 |  | $\begin{aligned} & 0 \\ & 43646 \end{aligned}$ | 4481 |
|  | 0 | 0 | 2988 | 18438 | 15167 | 27342 | $13892$ | $18263$ |  |  |
|  | 1695 | 324 | 387 | 308 | 2762 |  |  |  |  |  |
| 2000 | 1 | 2 | 0 | 0 | 1 | $\begin{aligned} & -1 \\ & 3269 \end{aligned}$ |  |  |  | 7075 |
|  | 0 | 15 | 60 | 2475 | 7585 |  |  |  | 829 |  |
|  | 633 | 174 | 39 | 96 | 421 |  |  |  |  |  |
| 2001 | 1 | 2 | 0 | 0 | 1 |  | $\begin{aligned} & -1 \\ & 6925 \end{aligned}$ | $\begin{aligned} & 50 \\ & 5181 \end{aligned}$ |  | 0 |
|  | 0 | 0 | 176 | 884 | 19449 |  |  |  |  | 2797 |
|  | 9505 | 843 | 625 | 121 | 258 |  |  |  |  |  |
| 2002 | 1 | 2 | 0 | 0 |  |  |  | 50 | 0 | 0 |
|  | 0 | 3 | 37 | 2349 | 1558 | $23776$ | 9568 | 6215 | 5901 | 444 |
|  |  |  | 0 | 0 |  |  |  |  |  |  |
| 2003 | 1 | 2 |  | 09885 | $36543$ | $\begin{aligned} & -1 \\ & 7420 \end{aligned}$ | $\begin{aligned} & -1 \\ & 37748 \end{aligned}$ | $\begin{aligned} & 50 \\ & 9582 \end{aligned}$ | $\begin{aligned} & 0 \\ & 2943 \end{aligned}$ | 03029 |
|  | 0 |  |  |  |  |  |  |  |  |  |
|  | 576 |  |  |  |  |  |  |  |  |  |
| 20042005 | 1 |  | 0 | $\begin{aligned} & 13054 \\ & 1011 \end{aligned}$ | 115006 | $\begin{aligned} & -1 \\ & 50232 \end{aligned}$ | $\begin{aligned} & -1 \\ & 3340 \end{aligned}$ | $\begin{aligned} & 50 \\ & 21608 \end{aligned}$ | $\begin{aligned} & 0 \\ & 8363 \end{aligned}$ | $\begin{aligned} & 0 \\ & 369 \end{aligned}$ |
|  | 0 |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | 20 |  |  |  |  |  |
|  |  |  |  | 0 | 1 | -1 | -1 | 50 | 0 | 0 |
|  |  |  | 30 | 2404 | 17514 | 16457 | 19899 | 2180 | 5924 | 0 |
|  | 2093 | 113 | 0 | 45 | 686 |  |  |  |  |  |
|  |  |  | 0 | 0 | 1 | -1 | -1 | 50 | 0 | 0 |
|  |  |  | 105 | 263 | 282 | 1892 | 660 | 1432 | 118 | 82 |
|  |  | 0 | 0 | 0 | 26 |  |  |  |  |  |
|  | 1 | 2 | 0 | 0 | 1 | -1 | -1 |  |  |  |
|  | 0 | 0 | 659 | 4305 | 12037 | 9213 | 11685 | 4780 | 3249 | 1079 |
|  | 1380 | 21 | 64 | 32 | 207 |  |  |  |  |  |
| 2008 | 1 | 2 | 0 | 0 | 1 | -1 | -1 | 50 | 0 | 0 |
|  | 0 | 52 | 513 | 1775 | 3779 | 2060 | 1627 | 1794 | 870 | 1106 |
|  | 35 | 211 | 565 | 0 | 47 |  |  |  |  |  |
| 2009 | 1 | 2 | 0 | 0 | 1 | -1 | -1 | 50 | 0 | 0 |
|  | 0 | 0 | 101 | 712 | 2439 |  | 945 | 880 | 189 | 334 |
|  | 194 | 12 | 190 | 0 | 1 |  |  |  |  |  |
| 2010 | 1 | 2 | 0 | 0 | 1 | -1 | -1 | 50 | 0 | 0 |
|  | 0 | 9 | 36 | 1741 | 5545 | 8261 | 6677 | 4755 | 403 | 3786 |
|  | 152 | 294 | 313 | 551 | 51 |  |  |  |  |  |


| 2011 | 1 | 2 | 0 | 0 | 1 | -1 | -1 | 50 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 0 | 0 | 255 | 4397 | 10231 | 13639 | 15908 | 13641 | 4424 | 4232 |
|  | 2773 | 1688 | 1003 | 264 | 424 |  |  |  |  |  |
| 2012 | 1 | 2 | 0 | 0 | 1 | -1 | -1 | 50 | 0 | 0 |
|  | 0 | 0 | 391 | 4456 | 10762 | 10003 | 8746 | 5782 | 2738 | 1133 |
|  | 289 | 433 | 143 | 127 | 226 |  |  |  |  |  |

0 \#_N_MeanSize-at-Age_obs
\#Yr Seas Flt/Svy Gender Part Ageerr Ignore datavector(female-male)
\# samplesize(female-male)

0 \#_N_environ_variables
0 \#_N_environ_obs
0 \# N sizefreq methods to read
0 \# no tag data

0 \# no morphcomp data
999 \# end of data file marker

## Appendix 3

Content of Stock Synthesis Starter file (Starter.SS) used at WGCSE 2014.
\#V3.23b
\#C Bass initial assessment
BassIVVII.dat
BassIVVII.ctl
0 \# 0=use init values in control file; 1=use ss3.par
1 \# run display detail ( $0,1,2$ )
1 \# detailed age-structured reports in REPORT.SSO $(0,1)$
0 \# write detailed checkup.sso file (0,1)
4 \# write parm values to ParmTrace.sso ( $0=$ no, $1=$ good,active; $2=$ good, all; $3=$ every iter,all_parms; $4=$ every,active)

1 \# write to cumreport.sso ( $0=$ no, $1=$ like\&timeseries; $2=$ add survey fits)
1 \# Include prior_like for non-estimated parameters $(0,1)$
1 \# Use Soft Boundaries to aid convergence $(0,1)$ (recommended)
2 \# Number of bootstrap datafiles to produce
8 \# Turn off estimation for parameters entering after this phase
10 \# MCMC burn interval
2 \# MCMC thin interval
0 \# jitter initial parm yalue by this fraction
-1 \# min yr for sdreport outputs (-1 for styr)
-1 \# max yr for sdreport outputs (-1 for endyr; -2 for endyr+Nforecastyrs
0 \# Nindividual STD years
\#vector of year values
0.0001 \#0.0001 \# final convergence criteria (e.g. 1.0e-04)

0 \# retrospective year relative to end year (e.g. -4)
\# min age for calc of summary biomass
2 \# Depletion basis: denom is: $0=$ skip; $1=$ rel X*B0; $2=$ rel X*Bmsy; 3=rel X*B_styr
0.4 \# Fraction (X) for Depletion denominator (e.g. 0.4)

2 \# SPR_report_basis: $0=$ skip; $1=(1-S P R) /\left(1-S P R \_t g t\right) ; 2=(1-S P R) /\left(1-S P R \_M S Y\right) ; 3=(1-S P R) /(1-$ SPR_Btarget); 4=rawSPR

2 \# F_report_units: $0=$ skip; 1=exploitation(Bio); 2=exploitation(Num); 3=sum(Frates)
\#COND 310 \#_min and max age over which average F will be calculated with F_reporting=4
0 \# F_report_basis: $0=$ raw; $1=F / F_{\text {spr; }} 2=F / F_{\text {MSY }} ; 3=F / F_{b t g t}$
999 \# check value for end of file

## Appendix 4

Content of Stock Synthesis Forecast file (Forecast.SS) used at WGCSE 2014. This is not used for creating forecasts yet, but has to be available.
\#V3.23b
\# for all year entries except rebuilder; enter either: actual year, -999 for styr, 0 for endyr, neg number for rel. endyr

1 \# Benchmarks: 0=skip; 1=calc F_spr,F_btgt,F_msy
1 \# MSY: 1= set to F(SPR); 2=calc F(MSY); 3=set to F(Btgt); 4=set to F(endyr)
0.4 \# SPR target (e.g. 0.40)
0.4 \# Biomass target (e.g. 0.40)
\#_Bmark_years: beg_bio, end_bio, beg_selex, end_selex, beg_relf, end_relf (enter actual year, or values of 0 or -integer to be rel. endyr)

000000
\# 201020102010201020102010 \# after processing
1 \#Bmark_relF_Basis: 1 = use year range; 2 = set relF same as forecast below
\#
0 \# Forecast: $0=$ none; $1=\mathrm{F}(\mathrm{SPR}) ; 2=\mathrm{F}(\mathrm{MSY}) 3=\mathrm{F}$ (Btgt); 4=Ave F (uses first-last relF yrs); $5=$ input annual F scalar

3 \# N forecast years
1 \# F scalar (only used for Do_Forecast-
\#_Fcast_years: beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -

\# 11808905/5 166759281 76317130 \# after processing
1 \# Control rule method ( $1=$ catch $=\mathrm{f}(\mathrm{SSB}$ ) west coast; $2=\mathrm{F}=\mathrm{f}(\mathrm{SSB})$ )
0.4 \#Control rule Biomass level for constant F (as frac of Bzero, e.g. 0.40 ); (Must be > the no F level below)

1 \# Control rule Biomass level for no F (as frac of Bzero, e.g. 0.10)
0.75 \# Control rule target as fraction of Flimit (e.g. 0.75)

3 \#_N forecast loops ( $1=$ OFL only; $2=\mathrm{ABC} ; 3=$ get F from forecast ABC catch with allocations applied)

3 \#_First forecast loop with stochastic recruitment
0 \#_Forecast loop control \#3 (reserved for future bells\&whistles)
0 \#_Forecast loop control \#4 (reserved for future bells\&whistles)
0 \#_Forecast loop control \#5 (reserved for future bells\&whistles)
2011 \#FirstYear for caps and allocations (should be after years with fixed inputs)
0 \# stddev of $\log$ (realized catch/target catch) in forecast (set value>0.0 to cause active impl_error)

0 \# Do West Coast gfish rebuilder output (0/1)
-1 \# Rebuilder: first year catch could have been set to zero (Ydecl)(-1 to set to 1999)
-1 \# Rebuilder: year for current age structure (Yinit) ( -1 to set to endyear+1)
1 \# fleet relative F: 1=use first-last alloc year; 2=read seas(row) x fleet(col) below
\# Note that fleet allocation is used directly as average F if Do_Forecast=4
0 \# basis for fcast catch tuning and for fcast catch caps and allocation (2=deadbio; 3=retainbio; 5=deadnum; 6=retainnum)

\# fleet assignment to allocation group (enter group ID\# for each fleet, 0 for not included in an alloc group)

0
\#_Conditional on $>1$ allocation group
\# allocation fraction for each of: 0 allocation groups
\# no allocation groups
0 \# Number of forecast catch levels to input (else calc catch from forecast F)

\#
999 \# verify end of input


[^0]:    ${ }^{1}$ Region 1: All waters which lie to the north and west of a line running from a point at latitude $48^{\circ} \mathrm{N}$, longitude $18^{\circ} \mathrm{W}$; thence due north to latitude $60^{\circ} \mathrm{N}$; thence due east to longitude $5^{\circ} \mathrm{W}$; thence due north to latitude $60^{\circ} 30^{\prime} \mathrm{N}$; thence due east to longitude $4^{\circ} \mathrm{W}$; thence due north to latitude $64^{\circ} \mathrm{N}$; thence due east to the coast of Norway.

