# WORKING GROUP FOR THE CELTIC SEAS ECOREGION (WGCSE) 

## VOLUME 1 | ISSUE 29

ICES SCIENTIFIC REPORTS

RAPPORTS
SCIENTIFIQUES DU CIEM


[^0]
## International Council for the Exploration of the Sea Conseil International pour l'Exploration de la Mer

H.C. Andersens Boulevard 44-46

DK-1553 Copenhagen V
Denmark
Telephone (+45) 33386700
Telefax (+45) 33934215
www.ices.dk
info@ices.dk

The material in this report may be reused for non-commercial purposes using the recommended citation. ICES may only grant usage rights of information, data, images, graphs, etc. of which it has ownership. For other third-party material cited in this report, you must contact the original copyright holder for permission. For citation of datasets or use of data to be included in other databases, please refer to the latest ICES data policy on ICES website. All extracts must be acknowledged. For other reproduction requests please contact the General Secretary.

This document is the product of an expert group under the auspices of the International Council for the Exploration of the Sea and does not necessarily represent the view of the Council.

## ICES Scientific Reports

## Volume 1 | Issue 29

# WORKING GROUP FOR THE CELTIC SEAS ECOREGION (WGCSE) 

Recommended format for purpose of citation:

ICES. 2019. Working Group for the Celtic Seas Ecoregion (WGCSE).
ICES Scientific Reports. 1:29. 1604 pp. http://doi.org/10.17895/ices.pub. 4982

## Editors

Timothy Earl • Sofie Nimmegeers


#### Abstract

Authors

Mikel Aristegui • Ewen Bell • Lynda Blackadder •Katie Boyle •Rui Catarino • Paul Coleman • Helen Dobby $\bullet$ Jennifer Doyle • Mickael Drogou • Timothy Earl • Simon Fischer • Helen Holah • Kieran Hyder • Andrzej Jaworski • Vladimir Klivnoi • Vladimir Laptikovsky • Mathieu Lundy • Claire Moore • Sara-Jane Moore • Sofie Nimmegeers • Lisa Readdy • Marianne Robert • Marta Quinzan • Stephen Shaw • David Stokes • Pia Schuchert • Katie Thomas • Bart Vanelslander • Jonathan White


## Contents

i Executive summary ..... xvi
ii Expert group information ..... xvii
1 Introduction ..... iv
2 Anglerfish (Lophius budegassa, Lophius piscatorius) in subareas 4 and 6 and in Division 3.a (North Sea, Rockall and West of Scotland, Skagerrak and Kattegat) ..... 1
2.1 General ..... 1
Stock description and management units ..... 1
Management applicable to 2018 and 2019 ..... 2
2.2 Data ..... 7
Landings ..... 7
Discards ..... 8
Biological ..... 8
Research vessel surveys ..... 8
Commercial catch-effort data ..... 9
2.3 Historical stock development ..... 10
2.4 Short-term projections ..... 10
2.5 Biological reference points ..... 10
Yield-per-recruit analysis and harvest rates ..... 11
2.6 Management plans ..... 11
2.7 Uncertainties and bias in assessment and forecast ..... 11
Commercial data ..... 11
Survey data ..... 12
Biological information ..... 12
Stock structure ..... 12
2.8 Recommendations for next Benchmark ..... 13
2.9 Management considerations ..... 13
2.10 References ..... 13
3 Cod in Division 6.a ..... 40
3.1 Introduction ..... 40
3.2 General ..... 40
3.2.1 Advice ..... 40
ICES Advice applicable for 2018 and 2019 ..... 40
ICES Advice applicable for 2015 ..... 40
3.2.2 Stock definition and the management unit ..... 40
Recent management ..... 41
TAC for 2012-2014 ..... 41
TAC for 2015-2018 ..... 41
TAC 2019 ..... 42
3.2.3 The fishery in 2018 ..... 42
3.3 Data ..... 43
Catch data ..... 43
Age compositions ..... 44
Weight-at-age ..... 44
Survey data ..... 44
Biological data ..... 45
3.4 Stock assessment ..... 46
Data screening ..... 46
Final assessment ..... 47
Stock status ..... 48
3.5 Short-term stock projections ..... 49
3.5.1 Reference points ..... 50
3.5.2 Management plans ..... 50
3.6 Uncertainties and bias in assessment and forecast ..... 51
Landings ..... 51
Discards ..... 51
Biological factors ..... 51
Stock structure ..... 52
Assessment method ..... 52
3.6.1 Recommendation for next Benchmark ..... 52
3.6.2 Management considerations ..... 53
3.6.3 Frequency of assessment ..... 53
3.6.4 References ..... 53
$4 \quad$ Cod (Gadus morhua) in Division 6.b (Rockall) ..... 110
4.1 General ..... 110
4.2 Data. ..... 111
4.3 References ..... 112
5 Cod in 7.a (Irish Sea) ..... 119
5.1 General ..... 119
5.2 Historical stock development ..... 122
5.3 Short-term predictions ..... 124
5.4 Biological reference points ..... 125
5.5 Management plans ..... 125
5.6 Uncertainties and bias in assessment ..... 125
5.7 Management considerations ..... 126
5.8 Future Issues and considerations ..... 126
5.9 References ..... 126
$6 \quad$ Cod in Divisions 7.e-k (Eastern English Channel and southern Celtic Seas) ..... 159
6.1 General. ..... 159
6.2 Data ..... 162
6.3 Stock assessment ..... 164
6.4 Short-term projections ..... 165
6.5 Medium-term projection ..... 166
6.6 Biological reference points ..... 166
6.7 Management plans ..... 167
6.8 Management considerations ..... 167
6.9 References ..... 168
$7 \quad$ Haddock (Melanogrammus aeglefinus) in Division 6.b (Rockall) ..... 214
Type of assessment in 2019: Update assessment taking into account the recommendations of benchmark 2019 ..... 214
ICES advice applicable to 2019 ..... 214
7.1 General. ..... 214
Stock description and management units ..... 214
Management applicable to 2018 and 2019 ..... 214
Fishery in 2018 ..... 216
Russian fishery in 2018 ..... 216
Scottish fishery in 2018 ..... 216
Irish fishery in 2018 ..... 216
Norwegian fishery in 2018. ..... 216
7.2 Data ..... 216
Landings ..... 216
BMS landings ..... 216
Discards ..... 217
Biological ..... 217
Surveys ..... 217
Commercial Effort, Lpue and Cpue. ..... 218
7.3 Description of stock assessment approach. ..... 218
Data screening. ..... 219
Final update assessment ..... 220
Final run ..... 220
Comparison with previous assessments ..... 220
State of the stock ..... 220
Statistical catch-at-age analysis (SCAA) ..... 221
Results of stock assessment by SAM model (state-space assessment model) ..... 221
The comparison of stock assessment results produced by different models ..... 221
7.4 Short-term projections ..... 221
Estimating year-class abundance ..... 221
Catch Constraint ..... 222
Mean Weights and F pattern ..... 222
Partitioning of catch into discards and landings ..... 222
STF results ..... 222
7.5 MSY evaluations and Biological reference points ..... 222
7.6 Management plans ..... 223
7.7 Uncertainties and bias in assessment and forecast ..... 224
7.8 Recommendation for next benchmark ..... 224
7.9 Management considerations ..... 225
7.10 References ..... 225
8 Haddock (Melanogrammus aeglefinus) in Division 7.a (Irish Sea) ..... 275
8.1 General ..... 275
8.2 Data ..... 277
8.3 Assessment ..... 279
8.4 Short-term projections ..... 282
8.5 Biological reference points ..... 282
8.6 Management plans ..... 283
8.7 Uncertainties and bias in assessment and forecast ..... 283
8.8 Recommendations for next benchmark assessment ..... 284
8.9 References ..... 284
9 Haddock in divisions 7.b,c,e-k ..... 309
9.1 General ..... 309
9.2 The fishery ..... 310
9.2.1 Information from the Industry ..... 311
9.3 Data ..... 311
9.3.1 Landings and discard numbers-at-age ..... 311
9.3.2 Biological. ..... 312
9.3.3 Surveys and commercial tuning fleets ..... 312
9.4 Historical stock development ..... 312
9.4.1 Data screening ..... 312
9.4.2 Final update assessment ..... 312
9.4.3 Comparison with previous assessments ..... 313
9.4.4 State of the stock ..... 313
9.5 Short-term projections ..... 313
9.6 MSY evaluations and biological reference points ..... 314
9.7 Management plans ..... 315
9.8 Uncertainties and bias in assessment and forecast ..... 315
9.8.1 Landings ..... 315
9.8.2 Discards ..... 316
9.8.3 Selectivity ..... 316
9.8.4 Assessment bias ..... 317
9.9 Surveys ..... 317
9.9.1 Forecast ..... 317
9.10 Recommendation for next benchmark ..... 318
9.10.1 Stock audit ..... 318
9.10.2 Recommendations for future work ..... 318
9.11 Management considerations ..... 318
9.12 References ..... 319
10 Megrim (Lepidorhombus ssp.) in divisions 4.a and 6.a (northern North Sea, West of Scotland) ..... 355
10.1 General ..... 355
10.2 Estimation of survey cpue indices ..... 358
10.3 Stock assessment ..... 359
10.4 Short-term projections ..... 359
10.5 Biological reference points ..... 360
10.6 References ..... 361
11 Megrim (Lepidorhombus spp.) in Division 6.b (Rockall) ..... 382
11.1 Uncertainties and bias in assessment and forecast ..... 384
11.2 Recommendation for next Benchmark ..... 385
11.3 References ..... 385
12 Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 11 (West of Scotland, North Minch) ..... 392
Type of assessment in 2019 ..... 392
ICES advice applicable to 2018 ..... 392
ICES advice applicable to 2019 ..... 392
12.1 General. ..... 392
Stock description and management units ..... 393
Management applicable to 2018 and 2019 ..... 393
Ecosystem aspects ..... 393
Fishery description ..... 393
12.2 Data available ..... 394
12.3 Assessment ..... 396
12.4 Catch option table ..... 397
12.5 Reference points ..... 397
12.6 Management strategies ..... 398
12.7 Quality of assessment and forecast ..... 398
12.8 Recommendation for next benchmark ..... 399
12.9 Management considerations ..... 399
12.10 References ..... 400
13 Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 12 (West of Scotland, South Minch) ..... 426
Type of assessment in 2019 ..... 426
ICES advice applicable to 2018 ..... 426
ICES advice applicable to 2019 ..... 426
13.1 General ..... 426
Stock description ..... 426
Management applicable to 2018 and 2019 ..... 426
Ecosystem aspects ..... 426
Fishery description ..... 427
13.2 Data available ..... 427
InterCatch ..... 427
Commercial catch ..... 427
Effort data ..... 428
Sampling levels ..... 428
Length compositions ..... 428
Sex ratio ..... 428
Mean weights ..... 429
Discarding ..... 429
Abundance indices from UWTV surveys ..... 429
13.3 Assessment ..... 430
Comparison with previous assessments ..... 430
State of the stock ..... 430
13.4 Catch scenarios table ..... 430
13.5 Reference points ..... 431
13.6 Management strategies ..... 431
13.7 Quality of assessment and forecast ..... 432
13.8 Recommendation for next benchmark ..... 433
13.9 Management considerations ..... 433
13.10 References ..... 433
14 Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 13 (West of Scotland, the Firth of Clyde and Sound of Jura) ..... 449
Type of assessment in 2019 ..... 449
ICES advice applicable to 2018 ..... 449
ICES advice applicable to 2019 ..... 449
14.1 General ..... 449
Stock description ..... 449
Management applicable to 2018 and 2019 ..... 450
Ecosystem aspects ..... 450
Fishery description ..... 450
14.2 Data available ..... 451
InterCatch ..... 451
Commercial catch ..... 451
Effort data ..... 451
Sampling levels ..... 451
Length compositions ..... 452
Sex ratio ..... 452
Mean weights ..... 452
Discarding ..... 453
Abundance indices from UWTV surveys ..... 453
14.3 Assessment ..... 454
Comparison with previous assessments ..... 454
State of the stock ..... 454
14.4 Catch option table ..... 454
14.5 Reference points ..... 455
14.6 Management strategies ..... 456
14.7 Quality of assessment and forecast ..... 456
14.8 Recommendation for next benchmark ..... 457
14.9 Management considerations ..... 457
14.10 References ..... 458
15 Norway lobster (Nephrops norvegicus) in Division 7.a, Functional Unit 14 (Irish Sea East) ..... 478
15.1 Nephrops Subarea 7 general section ..... 478
Stock description and management units ..... 478
Landings Obligation ..... 479
Minimum Conservation Reference Size (Minimum landing size) ..... 479
Management applicable in 2017 and 2018 ..... 479
Nephrops FU14 section ..... 485
Type of assessment in 2019 ..... 485
ICES advice applicable to 2018 ..... 485
ICES advice applicable to 2019 ..... 485
15.2 General ..... 485
Stock description and management units ..... 485
Fishery in 2018 ..... 486
Information from stakeholders ..... 486
15.3 Data ..... 487
InterCatch ..... 487
Landings ..... 487
Effort ..... 487
Sampling levels ..... 487
Commercial length-frequency distributions ..... 487
Length composition ..... 488
Sex ratio ..... 488
Mean weight explorations ..... 488
Discarding ..... 488
Abundance indices from UWTV surveys ..... 489
15.4 Assessment ..... 489
Comparison with previous assessments ..... 489
State of the stock ..... 490
15.5 Catch scenarios table ..... 490
15.6 Reference points ..... 490
15.7 Management strategies ..... 491
15.8 Quality of assessment and forecast ..... 491
15.9 Recommendation for next benchmark ..... 492
15.10 Management considerations ..... 492
15.11 References ..... 492
16 Norway lobster (Nephrops norvegicus)in Division 7.a, Functional Unit 15 (Irish Sea, west) ..... 509
Type of assessment ..... 509
ICES advice applicable to 2018 ..... 509
ICES advice applicable to 2019 ..... 509
16.1 General ..... 509
Stock description and management units ..... 509
Fishery description ..... 509
The fishery in 2017 ..... 510
Information from stakeholders ..... 510
16.2 Data ..... 510
InterCatch ..... 510
Landings ..... 510
Effort ..... 510
Sampling levels ..... 511
Commercial length-frequency distributions ..... 511
Sex ratio ..... 511
Mean weights ..... 511
Discards ..... 512
Surveys ..... 512
Abundance indices from UWTV surveys ..... 512
Nephrops trawl surveys ..... 512
16.3 Assessment ..... 513
Comparison with previous assessments ..... 513
State of the stock ..... 513
16.4 Catch option table ..... 513
16.5 Reference points ..... 514
16.6 Management strategy ..... 514
16.7 Quality of assessment and forecast ..... 514
16.8 Recommendations for next benchmark ..... 515
16.9 Management considerations ..... 515
16.10 References ..... 516
17 Norway lobster (Nephrops norvegicus) in divisions 7.b-c and 7.j-k, Functional Unit 16 (west and southwest of Ireland, Porcupine Bank) ..... 533
Type of assessment in 2019 ..... 533
ICES advice applicable to 2018 ..... 533
ICES advice applicable to 2019 ..... 533
17.1 General ..... 533
Stock description and management units ..... 533
Management applicable to 2018 and 2019 ..... 534
TAC in 2018 ..... 534
TAC in 2019 ..... 535
17.2 Closed area restrictions ..... 536
17.3 Fishery in 2018 and 2019 ..... 537
Effect of regulations ..... 537
Information from stakeholders ..... 537
17.4 Data. ..... 538
InterCatch ..... 538
Landings ..... 538
Sampling levels ..... 538
Commercial length-frequency distributions ..... 538
Sex ratio ..... 538
Mean weight explorations ..... 539
Discards ..... 539
Abundance indices from UWTV surveys ..... 539
Trawl surveys ..... 539
Commercial CPUE ..... 540
17.5 Stock assessment ..... 540
Comparison with previous assessments ..... 540
State of the stock ..... 540
Catch options table ..... 541
17.6 Reference points ..... 541
17.7 Management strategies ..... 541
17.8 Quality of assessment and forecast ..... 541
17.9 Recommendation for next benchmark ..... 542
17.10 Management considerations ..... 542
17.11 References ..... 543
18 Norway lobster (Nephrops norvegicus) in Division 7.b, Functional Unit 17 (west of Ireland, Aran Grounds) ..... 562
Type of assessment in 2019 ..... 562
ICES advice applicable to 2018 ..... 562
ICES advice applicable to 2019 ..... 562
18.1 General ..... 562
Stock description and management units ..... 562
Ecosystem aspects ..... 563
Fishery description ..... 563
Fishery in 2018 ..... 563
Information from stakeholders ..... 563
18.2 Data ..... 564
InterCatch ..... 564
Landings ..... 564
Effort ..... 564
Sampling levels ..... 564
Commercial length-frequency distributions ..... 564
Sex ratio ..... 564
Mean weight explorations ..... 564
Discarding ..... 565
Abundance indices from UWTV surveys ..... 565
18.3 Assessment ..... 565
Comparison with previous assessments ..... 565
State of the stock ..... 565
18.4 Catch scenario table ..... 566
18.5 Reference points ..... 566
18.6 Management strategies ..... 566
18.7 Quality of assessment and forecast ..... 567
18.8 Recommendation for next benchmark ..... 567
18.9 Management considerations ..... 567
18.10 References ..... 568
19 Norway lobster (Nephrops norvegicus) in divisions 7.a, 7.g and 7.j, Functional Unit 19 (Irish Sea, Celtic Sea, eastern part of southwest of Ireland) ..... 591
Type of assessment in 2019 ..... 591
ICES advice applicable to 2018 ..... 591
ICES advice applicable to 2019 ..... 591
19.1 General ..... 591
Stock description and management units ..... 591
Ecosystem aspects ..... 592
Fishery description ..... 592
Fishery in 2018 ..... 592
Information from stakeholders ..... 592
19.2 Data. ..... 592
InterCatch ..... 592
Landings ..... 593
Effort. ..... 593
Sampling levels ..... 593
Commercial length-frequency distributions ..... 593
Sex ratio ..... 594
Mean weight explorations. ..... 594
Discarding ..... 594
Abundance indices from UWTV surveys ..... 594
Information from Irish Groundfish survey. ..... 595
19.3 Assessment ..... 595
Comparison with previous assessments ..... 595
State of the stock ..... 596
19.4 Catch scenario table ..... 596
19.5 Reference points ..... 596
19.6 Management strategies ..... 597
19.7 Quality of assessment and forecast ..... 597
19.8 Recommendations for next benchmark ..... 598
19.9 Management considerations ..... 598
19.10 References ..... 598
20 Norway lobster (Nephrops norvegicus) in divisions 7.g and 7.h, Functional Units 20 and 21 (Celtic Sea) ..... 621
Type of assessment in 2019 ..... 621
ICES advice applicable to 2018 ..... 621
ICES advice applicable to 2019 ..... 621
20.1 General ..... 621
Stock description and management units ..... 621
Ecosystem aspects ..... 622
Fishery description ..... 622
Fishery in 2018 ..... 623
Ireland ..... 623
France ..... 623
UK ..... 623
Information from stakeholders ..... 623
20.2 Data ..... 623
InterCatch ..... 623
Landings ..... 623
Effort ..... 623
Sampling levels ..... 624
Commercial length-frequency distributions ..... 624
Sex ratio ..... 624
Mean weight explorations ..... 624
Discards ..... 625
Abundance indices from UWTV surveys ..... 625
Groundfish survey data ..... 626
20.3 Assessment ..... 626
Comparison with previous assessments ..... 626
State of the stock ..... 626
20.4 Catch scenario table ..... 627
20.5 Reference points ..... 627
20.6 Management plans ..... 628
20.7 Quality of assessment and forecast ..... 628
20.8 Recommendations for next benchmark ..... 629
20.9 Management considerations ..... 629
20.10 References ..... 630
21 Norway lobster (Nephrops norvegicus) in divisions 7.g and 7.f, Functional Unit 22 (Celtic Sea, Bristol Channel) ..... 650
Type of assessment in 2019 ..... 650
ICES advice applicable to 2018 ..... 650
ICES advice applicable to 2019 ..... 650
21.1 General ..... 650
Stock description and management units ..... 650
Ecosystem aspects ..... 651
Fishery description ..... 651
Fishery in 2018 ..... 651
Information from stakeholders ..... 652
21.2 Data ..... 652
InterCatch ..... 652
Landings ..... 652
Effort ..... 652
Sampling levels ..... 652
Sampling and Raising Procedure Review ..... 652
Commercial length-frequency distributions ..... 653
Sex ratio ..... 653
Mean weight explorations ..... 653
Discarding ..... 653
Surveys ..... 654
Abundance indices from UWTV surveys ..... 654
Groundfish survey data ..... 654
21.3 Assessment ..... 654
Comparison with previous assessments ..... 654
State of the stock ..... 655
21.4 Catch scenarios table ..... 655
21.5 Reference points ..... 656
21.6 Management strategies ..... 656
21.7 Quality of assessment and forecast ..... 656
21.8 Recommendation for next benchmark ..... 657
21.9 Management considerations ..... 657
21.10 References ..... 658
22 Plaice in Division 27.7.a (Irish Sea) ..... 676
22.1 General ..... 676
22.2 Data ..... 678
22.3 Historical stock development ..... 680
22.4 Short-term projections ..... 683
22.5 Medium-term projections. ..... 683
22.6 MSY explorations ..... 684
22.7 Management plans ..... 684
22.8 Uncertainties and bias in assessment and forecast ..... 684
22.9 Recommendations for next benchmark ..... 684
22.10 Management considerations ..... 685
22.11 References ..... 685
23 Plaice (Pleuronectes platessa) in divisions 7.b-c (West of Ireland) ..... 712
Type of assessment in 2019 ..... 712
23.1 General. ..... 712
Stock Identity ..... 712
Data ..... 712
24 Plaice (Pleuronectes platessa) in Division 7.e (western English Channel ..... 716
24.1 General. ..... 716
24.2 Data ..... 718
24.3 Data-limited methods ..... 720
24.4 Stock assessment ..... 722
24.5 Short-term projections ..... 725
24.6 Biological reference points ..... 726
24.7 Exploratory assessment including discards ..... 727
24.8 Management plans ..... 727
24.9 Uncertainties and bias in assessment and forecast ..... 728
24.10 Recommendations for next Benchmark ..... 728
24.11 Management considerations ..... 729
24.12 References ..... 730
24.13 Tables ..... 731
24.14 Figures ..... 753
25 Plaice (Pleuronectes platessa) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea) ..... 772
25.1 Type of assessment in 2019 ..... 772
25.2 Data ..... 773
25.3 Stock assessment ..... 776
25.4 Short-term projections ..... 778
25.5 Precautionary approach reference points ..... 778
25.6 Management plans ..... 778
25.7 Uncertainties in assessment and forecast ..... 778
25.8 References ..... 779
26 Plaice (Pleuronectes platessa) in divisions 7h-k (Celtic Sea South, southwest of Ireland) ..... 813
26.1 General ..... 813
26.2 Data ..... 814
26.3 Historical stock development ..... 815
26.4 MSY evaluation ..... 817
26.5 Uncertainties and bias in the assessment and forecast ..... 818
26.6 Recommendations for the next benchmark ..... 819
26.7 Management considerations ..... 819
26.8 References ..... 820
27 Pollack (Pollachius pollachius) in subareas 6-7 (Celtic Seas and the English Channel) ..... 843
27.1 General ..... 843
27.1.1 Stock Identity ..... 843
27.1.2 Management applicable to 2019 ..... 843
27.2 Stock assessment ..... 845
27.3 Management considerations ..... 846
27.3.1 Management plan ..... 846
27.3.2 Recommendations ..... 846
27.4 References ..... 847
28 Saithe (Pollachius virens) in subareas 7-10 ..... 856
28.1 General ..... 856
28.1.1 Stock Identity ..... 856
28.1.2 Management applicable to 2019 ..... 856
28.2 Stock assessment ..... 858
28.2.1 Uncertainties in assessment and forecast ..... 858
28.2.2 Recommendations ..... 858
28.3 References ..... 858
29 Seabass (Dicentrarchus labrax) in divisions 4.b-c, 7.a, and 7.d-h (central and southern North Sea, Irish Sea, English Channel, Bristol Channel, and Celtic Sea) ..... 862
29.1 General ..... 862
29.1.1 Stock definition and ecosystem aspects ..... 862
29.1.2 Management ..... 862
29.1.2.1 Management applicable to 2018 ..... 863
29.1.3 Fishery description ..... 864
29.1.3.1 Total landings (official) ..... 864
29.2 Data ..... 864
29.2.1 Commercial landings ..... 864
29.2.2 Details of the methodology used to calculate French and UK historical landings can be found in the Stock Annex. Commercial length compositions ..... 864
29.2.2.1 Sampling rates ..... 865
29.2.2.2 Length composition estimates ..... 865
29.2.3 Commercial age composition ..... 871
29.2.4 Commercial discards ..... 876
29.2.5 Recreational catches ..... 877
29.2.5.1 Recreational catches before management measures (1985-2014) ..... 877
29.2.5.2 Recreational catches after management measures (2015-present) ..... 881
29.2.5.3 Inclusion in the stock synthesis model ..... 882
29.2.6 Biological data ..... 883
29.2.6.1 Growth parameters ..... 883
29.2.6.2 Standard deviations of length-at-age ..... 883
29.2.6.3 Age error parameters for Stock Synthesis ..... 883
29.2.6.4 Weight-at-length ..... 884
29.2.6.5 Maturity-at-length ..... 884
29.2.6.6 Natural mortality ..... 884
29.2.6.7 Post-realease mortality ..... 885
29.2.7 Survey data used in assessment ..... 886
29.2.7.1 Pre-recruit surveys in UK ..... 886
29.2.7.2 Pre-recruit surveys in France ..... 886
29.2.7.3 Channel Groundfish survey CGFS ..... 886
29.2.8 Commercial landings per unit of effort ..... 888
29.2.9 Other relevant data ..... 888
29.3 Stock assessment ..... 888
29.3.1 Model structure and input data / parameters for update assessment ..... 888
29.3.1.1 Model structure ..... 888
29.3.1.2 Fleet definition ..... 888
29.3.1.3 Landed catches ..... 888
29.3.1.4 Abundance indices ..... 889
29.3.1.5 Fishery landings age composition data ..... 889
29.3.1.6 Fishery landings length composition data ..... 889
29.3.1.7 Model assumptions and parameters ..... 889
29.3.1.8 Incorporation of recreational fishery catch estimates ..... 889
29.3.1.9 Final update assessment: diagnostics ..... 890
29.3.2 Analytical retrospective analyses. ..... 891
29.3.3 Final update assessment: long-term trends ..... 891
29.3.4 Comparison with previous assessments ..... 891
29.3.5 The state of the stock ..... 892
29.4 Biological reference points ..... 892
29.5 Short-term predictions ..... 893
29.5.1 Recruiting year-class strength ..... 893
29.5.2 Numbers of fish in 2019 ..... 894
29.5.3 F-at-age vectors ..... 894
29.5.4 Weights-at-age ..... 894
29.5.5 Maturity ogive ..... 894
29.5.6 Detailed short-term forecast output at status quo $F$ ..... 895
29.5.7 Management options ..... 895
29.6 Uncertainties and bias in assessment and forecast ..... 895
29.6.1 Landings and discards data ..... 895
29.6.2 Fishery composition data ..... 896
29.6.3 Recreational fishery harvests ..... 896
29.6.4 Surveys ..... 897
29.6.5 Commercial Ipue indices ..... 897
29.6.6 Stock structure and migrations ..... 898
29.6.7 Biological parameters ..... 898
29.6.8 Intermediate year fishing mortality and catch levels for forecasts ..... 898
29.7 Recommendations ..... 899
29.7.1 Management considerations ..... 899
29.8 References ..... 901
30 Seabass (Dicentrarchus labrax) in divisions 6.a,7.b and 7.j (West of Scotland, West of Ireland, eastern part of southwest of Ireland) ..... 968
30.1 General ..... 968
30.2 Data ..... 969
30.3 Historical stock development ..... 970
30.4 Management plans ..... 970
30.5 Management considerations ..... 970
30.6 Data needs ..... 970
30.7 References ..... 971
31 Sole in Division 7.a (Irish Sea) ..... 973
31.1 General ..... 973
31.2 Data ..... 975
31.3 Stock assessment ..... 977
31.4 Short-term projections ..... 979
31.5 MSY explorations ..... 980
31.6 Biological reference points ..... 981
31.7 Management Plans ..... 981
31.8 Uncertainties and bias in assessment and forecast ..... 981
31.9 Recommendations for next Benchmark ..... 982
31.10 Management considerations ..... 983
31.11 Ecosystem considerations ..... 983
31.12 References ..... 983
32 Sole (Solea solea) in divisions 7.b and 7.c (West of Ireland) ..... 1028
Type of assessment in 2019 ..... 1028
32.1 General ..... 1028
32.1.1 Stock identity ..... 1028
32.1.2 Data ..... 1028
32.1.3 Historical stock development ..... 1028
33 Sole (Solea solea) in Division 7.e (western English Channel) ..... 1034
33.1 Genera ..... 1034
33.2 Data ..... 1036
33.3 Stock assessment ..... 1038
33.4 Short-term projections ..... 1040
33.5 Biological reference points ..... 1041
33.6 Management plan ..... 1042
33.7 Uncertainties in assessment and forecast ..... 1042
33.8 Recommendation for the next benchmark ..... 1043
33.9 Management considerations ..... 1043
33.10 Ecosystem considerations and changes in the environment ..... 1043
33.11 Regulations and their effects ..... 1043
33.12 References ..... 1044
$34 \quad$ Sole in divisions 7.f and 7.g (Bristol Channel, Celtic Sea) ..... 1100
34.1 General ..... 1101
34.2 Data ..... 1102
34.3 Stock assessment ..... 1105
34.4 Short term projections ..... 1107
34.5 MSY explorations ..... 1108
34.6 Biological reference points ..... 1108
34.7 Management plans ..... 1109
34.8 Uncertainties and bias in assessment and forecast ..... 1109
34.9 Recommendation for next Benchmark ..... 1110
34.10 Management considerations ..... 1111
34.11 Ecosystem considerations ..... 1111
34.12 References ..... 1111
35 Sole (Solea solea) in divisions 7.h-k (Celtic Sea South, southwest of Ireland) ..... 1159
35.1 General ..... 1159
35.2 Data ..... 1160
35.2.1 Landings and discards ..... 1160
35.2.2 Landings numbers-at-age ..... 1161
35.2.3 Biological ..... 1161
35.2.4 Surveys and commercial tuning fleets ..... 1161
35.2.5 Data quality ..... 1161
35.3 Historical stock assessment development ..... 1161
35.3.1 Exploratory assessment ..... 1162
35.3.2 Final assessment ..... 1162
35.3.3 State of the stock ..... 1163
35.4 MSY evaluation ..... 1163
35.4.1 MSY and Biological reference points ..... 1164
35.5 Uncertainties and bias in the assessment and forecast ..... 1164
35.6 Recommendations for the next benchmark ..... 1164
35.7 Management considerations ..... 1164
35.8 References ..... 1165
36 Whiting (Merlangius merlangus) in Division 6.a (West of Scotland) ..... 1191
36.1 General ..... 1191
36.2 Data ..... 1192
36.3 Historical stock development ..... 1195
36.4 Short-term projections ..... 1197
36.5 MSY explorations ..... 1197
36.6 MSY and Biological reference points ..... 1197
36.7 Management plans ..... 1198
36.8 Uncertainties and bias in the assessment and forecast. ..... 1198
36.9 Recommendation for next benchmark ..... 1198
36.10 Management considerations ..... 1199
36.11 References ..... 1199
37 Whiting (Merlangius merlangus) in Division 6.b (Rockall) ..... 1246
37.1 General ..... 1246
37.2 Data ..... 1247
37.3 Target category ..... 1247
37.4 Management considerations ..... 1248
38 Whiting (Merlangius merlangus) in Division 7.a (Irish Sea) ..... 1254
38.1 General ..... 1254
38.2 Information from the Industry ..... 1256
38.3 Data ..... 1256
38.4 Historical stock development ..... 1257
38.5 Short-term predictions ..... 1259
38.6 Medium-term projection ..... 1259
38.7 MSY evaluations and Biological Reference Points ..... 1259
38.8 Management plans ..... 1259
38.9 Uncertainties and bias in assessment and forecast ..... 1259
38.10 Recommendations for next benchmark assessment ..... 1260
38.11 Management considerations ..... 1260
38.12 References ..... 1261
39 Whiting (Merlangius merlangus) in divisions 7.b-c and 7.e-k (southern Celtic Seas and eastern English Channel) ..... 1313
39.1 General. ..... 1313
39.2 The fishery in 2018 ..... 1315
39.3 Data ..... 1315
39.4 Historical stock development ..... 1318
39.5 State of the stock ..... 1320
39.6 Short-term projections ..... 1320
39.7 MSY evaluations and Biological reference points ..... 1320
39.8 Management plans ..... 1321
39.9 Uncertainties and bias in assessment and forecast ..... 1321
39.10 Recommendation for next benchmark ..... 1322
39.11 Management considerations ..... 1322
39.12 References ..... 1323
Annex 1: List of participants ..... 1356
Annex 2: Stock Annexes ..... 1358
Annex 3: Working documents ..... 1361
Annex 4: Technical Minutes ..... 1492

## i Executive summary

In total the ICES Working Group for the Celtic Seas Ecoregion (WGCSE) is responsible for the provision of updated fisheries data, assessments and draft advice for 40 demersal fish and Nephrops stocks across ICES subareas 6 and 7 (with the distribution of megrim, seabass, anglerfish and saithe extending into other divisions). This includes twelve Nephrops stocks, five sole and plaice stocks, four cod and whiting stocks, three haddock stocks, two each of megrim and seabass, one anglerfish, one saithe and one pollack stock. As in previous years, advice for Nephrops, anglerfish and Rockall megrim is not issued until autumn to make use of the most up to date survey information. Advice for Rockall haddock is not issued until autumn to allow evaluation of reference points as part of a benchmark currently underway. For a number of other stocks (bss.27.6a7bj, cod.27.6b, nep.27.6aoutFU, nep.27.7outFU, ple.27.7bc, sai.27.7-10, sol.27.7bc, whg.27.6a, whg.27.6b), no new advice was provided this year. Advice on the remaining stocks was released on the 28th June.

Since the last Working Group meeting, two stocks have gone through a benchmark or Interbenchmark procedure; cod.27.6a, had. 27.6 b and sol.27.7fg, the results of which were presented to the group.

Update assessments were generally carried out according to the stock annexes (any deviations were detailed in the stock sections). Overall the stock status across the ecoregion is very similar to that presented last year. Of the 40 stocks assessed, 21 were fished below FMSY, eight stocks were fished above $\mathrm{F}_{\mathrm{msy}}$ and 11 stocks had unknown status relative to $\mathrm{F}_{\text {MSY; }} 21$ were above MSY $B_{\text {trigger, }}$ and eight were below MSY $B_{\text {trigger, }}$ with 11 unknown relative to $B_{\text {trigger. }}$

## ii Expert group information

| Expert group name | Working Group for the Celtic Seas Ecoregion (WGCSE) |
| :--- | :--- |
| Expert group cycle | Annual |
| Year cycle started | 2019 |
| Reporting year in cycle | $1 / 1$ |
| Chairs | Sofie Nimmegeers, Belgium |
| Meeting venue and dates | $8-17$ May, 2019, Ghent, Belgium (29 participants) |

## 1 Introduction

### 1.1 Terms of reference

### 1.1.1 Generic ToRs for Regional and Species Working Groups

2018/2/ACOM05 The following ToRs apply to: AFWG, HAWG, NWWG, NIPAG, WGWIDE, WGBAST, WGBFAS, WGNSSK, WGCSE, WGDEEP, WGBIE, WGEEL, WGEF, WGHANSA and WGNAS.

## The working group should focus on:

a) Consider and comment on Ecosystem and Fisheries overviews where available;
b) For the aim of providing input for the Fisheries Overviews, consider and comment for the fisheries relevant to the working group on:

1. descriptions of ecosystem impacts of fisheries;
2. descriptions of developments and recent changes to the fisheries;
3. mixed fisheries considerations; and
4. emerging issues of relevance for the management of the fisheries.
c) Conduct an assessment on the stock(s) to be addressed in 2019 using the method (analytical, forecast or trends indicators) as described in the stock annex and produce a brief report of the work carried out regarding the stock, summarising where the item is relevant:
5. Input data and examination of data quality;
6. Where misreporting of catches is significant, provide qualitative and where possible quantitative information and describe the methods used to obtain the information;
7. For relevant stocks (i.e. all stocks with catches in the NEAFC Regulatory Area) estimate the percentage of the total catch that has been taken in the NEAFC Regulatory Area in 2018;
8. Estimate MSY proxy reference points for the category 3 and 4 stocks;
9. The developments in spawning-stock biomass, total stock biomass, fishing mortality, catches (wanted and unwanted landings and discards) using the method described in the stock annex;
10. The state of the stocks against relevant reference points;
11. Catch scenarios for next year(s) for the stocks for which ICES has been requested to provide advice on fishing opportunities;
12. Historical and analytical performance of the assessment and catch options with a succinct description of quality issues with these. For the analytical performance of category 1 and 2 age-structured assessment, report the mean Mohn's rho (assessment retrospective (bias) analysis) values for $\mathrm{R}, \mathrm{SSB}$ and $F$. The WG report should include a plot of this retrospective analysis. The values should be calculated in accordance with the "Guidance for completing ToR viii) of the Generic ToRs for Regional and Species Working Groups - Retrospective bias in assessment" and reported using the ICES application for this purpose. Produce a first draft of the advice on the stocks under considerations according to ACOM guidelines.
d) Review progress on benchmark processes of relevance to the Expert Group;
e) Prepare the data calls for the next year update assessment and for planned data evaluation workshops;
f) Identify research needs of relevance for the work of the Expert Group.

Information of the stocks to be considered by each Expert Group is available here.

### 1.1.2 Specific ToRs

WGCSE - Working Group for the Celtic Seas Ecoregion
2018/2/ACOM13 The Working Group for the Celtic Seas Ecoregion (WGCSE), chaired by Timothy Earl, UK and Sofie Nimmegeers*, Belgium will meet in Ghent, Belgium, 8-17 May 2019 and by correspondence September / October 2019 to:
a) Address generic ToRs for Regional and Species Working Groups;
b) Report on reopened advice as appropriate.

The assessments will be carried out on the basis of the stock annex. The assessments must be available for audit on the first day of the meeting.

Material and data relevant for the meeting must be available to the group on the dates specified in the 2019 ICES data call.

WGCSE will report by 31 May 2019 for the attention of ACOM, and by 7 October 2019 for Nephrops stocks, anglerfish and megrim in Rockall. Concerning ToR b) the group will report on the ACOM guidelines on reopening procedure of the advice before 12 October and will report on reopened advice before 28 October.

### 1.2 Participation

The number of participants able to attend the Working Group for the full duration of the meeting has increased slightly in the last three years (Figures 1.2.1-1.2.3). The Working Group continues to welcome a number of members participating remotely, or for part of the meeting. As last year, six institutes were represented by full-time participants attending the meeting.


Figure 1.2.1. Numbers of WGCSE participants over time, and whether they were full-time or part-time (part-time includes working by correspondence).


Figure 1.2.2. Number of participants from each institute attending by year.


Figure 1.2.3. Number of participants attending the meeting by day. Note that no plenary session was held on Sunday 12th.

### 1.3 Methods

The type of final assessments presented at the WG are summarised as follows:
Category 1 age-based assessments and forecasts were conducted for bss.27.4bc7ad-h, cod.27.6.a, cod.27.7.e-k, had.27.6.b, had.27.7.a, had.27.7.b-k, ple.27.7.a, sol.27.7.a, sol.27.7.e, sol.27.7.fg, whg.27.6.a (no forecast presented as the advice published in 2018 is still valid), whg.27.7.a, whg.27.7.b-ce-k;

Category 1 Bayesian surplus production model for lez.27.4.a6.a;
Category 1: UWTV survey based assessments and advice were used for nep.fu.11, nep.fu.12, nep.fu.13, nep.fu.14, nep.fu.15, nep.fu.16, nep.fu.17, nep.fu.19, nep.fu. 2021 and nep.fu.22. Fisheries data were updated at the May meeting and survey data were updated in the autumn;

Category 3: Catch-at-age based assessments with caveats i.e. used for trends only and without forecasts for ple.27.7.e, ple.27.7.h-k and sol.27.7.h-k;

Category 3: SPICT used to provide biomass trend for ple.27.7fg;
Category 3: Analysis of the trends in survey data are used as the basis for advice for anf.27.3a46, cod 27.7.a and lez.27.6b;

Category 4: Depletion corrected average catch was used for pol.27.67 and pok.27.7-10;
Category 5 \& 6: No assessments were carried out in 2019 for bss.27.6bc7ad-h, cod.6b, nep.27.6aoutFU and nep.27.7outFU, ple.27.7bc, sol.27.7bc, whg. 27.6 b ; only landings statistics were updated.

For the stocks for which a full analytical assessment was possible, the WG typically used either Extended Survivor's Analysis (XSA), Time-Series Analysis (TSA), or Age-Structured Assessment Program (ASAP). These approaches and procedures for using them are discussed in further detail in the relevant stock annexes.

### 1.4 Data issues

Data were generally submitted in a timely fashion through the InterCatch database for landings and discards data, and through the accessions database for other sources of data. There were no general data issues, those that affect specific stocks are discussed in the relevant section.

### 1.5 Transparent Assessment Framework (TAF)

TAF is a new framework, currently in development, to organize all ICES stock assessments. Using a standard sequence of $R$ scripts, it makes the data, analysis, and results available online, and documents how the data were pre-processed. Among the key benefits of this structured and open approach are improved quality assurance and peer review of ICES stock assessments. Furthermore, a fully scripted TAF assessment is easy to update and rerun later, with a new year of data. A number of assessments are being scripted in standard TAF scripts. See http://taf.ices.dk for more information.

During the WGCSE 2019 meeting, the following stocks made progress transferring their assessments into TAF: anf.27.3a46, cod.27.6a, cod.27.6b, cod.27.7a, had.27.6b, had.27.7b-k, nep.fu.11, nep.fu.22, ple.27.7e, ple.27.7h-k, pol.27.67, sol.27.7e, sol.27.7fg, sol.27.7h-k, whg.27.6a, whg.27.6b, whg.27.7b-ce-k.

### 1.6 Internal auditing and external reviews

This year ICES reintroduced the external review process, establishing a Review Group (RG) with the responsibility of auditing all category 1 stocks ahead of the ADG. This review group constitutes of 20 students from the University of Maine (US) and is led by a PhD student (Mackenzie Mazur) and the faculty advisor (Yong Chen). No major errors were detected by this process, but the RG did provide some useful comments to be considered in next year's assessments.

In addition, as in previous years the WG carried out its own internal audit process using the standard ICES template. Given the workload of many of the scientists at WGCSE (sometimes with one scientist responsible for two or more stocks), many of the reports were not finalized until after the WG meeting. Audits were therefore typically carried out by correspondence after the WG and not completed for some stocks.

All stocks for which advice was provided in June and October 2019 were audited by the WG and audit reports were produced for most of these. Issues discovered during the audit process were corrected in the WG report.

### 1.7 Generic ToR e: WGCSE recommendations for stocks to be benchmarked

The following stocks are recommended by WGCSE to be benchmarked in 2020: cod.27.7e-k, had. $27.7 \mathrm{~b}-\mathrm{k}$ and whg.27.7bce-k (WKCELTIC); sol.27.7fg and sol.27.7h-k together with three stocks from WGNSSK, sol.27.4, sol.27.7d and tur 27.3a (WKFlatNSCS); cod.27.6a and whg.27.6a (WKWest).

In February 2019, the first Data Evaluation Workshop in preparation of WKCELTIC was established with the focus on streamlining data compilation procedures for fishery-dependent and survey data. This will improve the transparency and diagnostics surrounding commercial tuning fleets and surveys. The second Data Evaluation Workshop is scheduled for October 2019.

WKCELTIC will revise the assessment methods and diagnostics, given the potential for changes in selectivity in the commercial fishery. Also mixed fisheries and multispecies interactions will be investigated as well as environmental drivers that may impact the growth and recruitment of all three species.

For sol.27.7fg and sol.27.7h-k (WKFlatNSCS), the focus is to examine alternative assessment models to XSA (e.g. A4A, ASAP, SAM, CASAL), explore the impact of all available tuning fleets, reconsider available life-history and catch data.

During the benchmark for cod.27.6a, the issues associated with stock structure will be addressed. A possible merge between North Sea and 27.6a cod stocks or alternatively, a split of area 27.6 a in two areas North and South, will be evaluated. Furthermore, misreporting of Scottish data, fisheries selectivity pattern and alternative discard models will be explored. For whg.27.6a, the focus is to address the emergence of a trend in survey catchability.

WGCSE recommend that pol.27.67 and pok.27.7-10 should be benchmarked together in 2021. Currently, both stocks are categorized as category 4 data-limited and the DCAC method is applied to provide advice. As the DCAC method only uses long time-series of official landings, it may not reflect recent stock fluctuations or changes in the fisheries, smoothed by the length of the time-series. So new computations of DCAC are always very close to the previous year's results, even if recruitment or SSB highly fluctuate. Therefore, it is relevant to explore new assessment models. Furthermore, this is the first year advice was provided for the pok.27.7-10 stock.

WGCSE recommend that ple.27.7h-k and ple. 27.7 fg should be benchmarked in 2021. For those plaice stocks similar issues will be addressed as scheduled for the WKFlatNSCS in 2020.

Further details are given in the stock sections.
This year a new initiative has been approved for prioritizing the stocks that need to be benchmarked. The sum of the weighting scores $(1-5)$ for each of the five criteria will determine the urgency for a benchmark. Those criteria are related to the quality of the previous assessments, the opportunity to improve the assessment, the management importance, the perceived stock status and the time since the previous benchmark.

To have an overview of this information, an issue list is requested for every stock. Later in the year, this prioritization exercise will be completed.

### 1.8 Specific ToRs

### 1.8.1 c(ii): Estimation of MSY proxy reference points for category 3 and 4 stocks

The Terms of Reference contained a list of six stocks for which proxy reference points should be considered. The Working Group addressed this ToR as follows.

## Category 3 stocks

For four stocks (cod.27.7e-k, ple.27.7e, ple.27.7h-k and sol.27.7h-k) age-based assessments are performed, although only used as relative indicators of stock status. For these stocks, most of the reference points were estimated using the package EqSim, and the method of WKMSYREF4 at WGCSE 2017. The extra data available at this year's Working Group did not warrant recalculation of the reference points.

For ple. 27.7 fg , a SpiCT assessment using survey and lpue data, combined with a hind-cast of discards was used to estimate the stock status relative to reference points.

For lez.27.6b, a SpiCT assessment using survey data was used to estimate the stock status relative to reference points.

For anf.27.3a46, which was benchmarked in 2018, none of the DL approaches for estimating proxy reference points were entirely satisfactory.

The assessment of cod.27.7a was not of sufficient quality to be retained as a category 1 assessment. Therefore, the reference points previously defined for this stock are not considered appropriate for providing advice. As a result, the advice is based on the survey index as indicator of stock size and the category 3 precautionary approach.

## Category 4 stocks

For pol.27.67 and pok.27.7-10, no reference points are defined.

### 1.8.2 c(viii): Calculation of Mohn's Rho

Through this additional ToR, the Working Group was requested to report the assessment bias statistic Mohn's rho for each of the category 1 stocks. For the following stocks the Mohn's rho data were uploaded to the "Retro-bias-2019" SharePoint: cod.27.6a, cod.27.7a, cod.27.7e-k, had.27.7b-k, lez.27.4a6a, ple.27.7a, sol.27.7a, sol.27.7e, sol.27.7fg, whg.27.7a and whg.27.7b-ce-k. The assessments of Nephrops stocks do not revise the perception of previous years, and so there is no retrospective assessment.
The guidance on calculating Mohn's Rho seems unclear about whether the SSB for the intermediate year should be used for the calculation of rho in XSA and ASAP models. Some members considered that the SSB in 2019 was a consequence of the known numbers and catch in 2018 and should therefore be included, while others took the view that the SSB in 2019 depends on a 2019 recruitment assumption where the recruitment has some proportion mature. Furthermore, the SSB 2019 also depends on the assumed stock weights-at-age for 2019. In the latter case, the SSB 2019 is not directly derived from the assessment model and should therefore be excluded from the Mohn's rho calculation. For most stocks that use an XSA or ASAP model, the final year was set to 2018, and in all cases we've asked for this to be made clear in the report. For cod.27.6a, that uses a TSA model, the SSB 2019 is projected by the model and thus included.

In November 2019 a workshop is planned (WKFORBIAS) to quantify the extent and possible causes for retrospective bias. However, as an interim solution for 2019, ICES has provided guidance (based on a paper of Hurtado et al., 2015), that suggests downgrading stocks from category 1 , if the Mohn's Rho value of the SSB retro is outside the range -0.15 to 0.2 .

The retrospective biases in cod.27.7e-k, cod.27.7a and had. $27.7 \mathrm{~b}-\mathrm{k}$ were all outside the limits of the rule of thumb of Hurtado-Ferro (2015). The whg.27.7b-ce-k, assessment was also revised substantially in 2019. The choice of terminal year made no difference to which stocks had SSB retro values outside the acceptable range (in the case of whg. $27.7 \mathrm{~b}-\mathrm{ce}-\mathrm{k}$ the Mohn's rho with 2019 terminal year was not presented).

We dealt with these on a case-by-case basis.
For cod.27.7a, the Mohn's Rho value for SSB was 0.92 (ending in 2018). The working group considered several alternative models for catch selectivity, reduced age range, and alternative values for natural mortality, but was unable to identify a single source of the bias. As a result, advice was provided as a category 3 assessment, using the NI Groundfish Quarter 1 WIBTS Survey, as this was considered the most reliable indicator of exploitable biomass. The Working Group concluded that the fit of the model was so poor that the reference points based on it were unreliable and used a qualitative evaluation instead.

For cod.27.7e-k, the Mohn's Rho value for SSB was 0.48 (ending in 2018) and therefore the stock was moved from category 1 to category 3 (i.e. trends in the assessment). Unlike Cod.27.7a, the Working Group had enough confidence in the model outputs to determine the status relative to the reference points.

For whg. $27.7 \mathrm{~b}-\mathrm{ce}-\mathrm{k}$, the Mohn's Rho value for SSB was 0.06 (ending in 2018), but the retrospective on $R$ has a value of 0.881 due to a substantial revision downwards. The retrospective appears to be a result of very low catches in the combined EVHOE-WIBTS-Q4 (French) and IGFS-WIBTSQ4 (Irish) at ages 1-3, which is reflected in a low level of discards in the French OTB fleet, but not in the Irish OTB fleet (predominantly sampled from 7 b ). The consistency of the data suggests some movement of the smaller fish in the stock, possibly related to a slightly later timing of the survey. Given the multiple sources of data revising our perception, (French survey, Irish Survey, French discards) the Working Group considered that the perception of the stock has changed due to additional data, and it was appropriate for the model to revise recruitment and SSB down in recent years. The Category 1 approach was used.

For had.27.7b-k, the Mohn's Rho value for SSB was -0.17 , largely due to a change in the assessment in 2019, rather than systematic revisions. This was believed to be mainly due to revisions in stock weights in the most recent assessment. The Working Group concluded that the assessment was still suitable for providing category 1 advice.

Note that Celtic Sea cod/haddock/whiting are undergoing a benchmark in 2020, which is likely to result in a revised perception of the stocks next year.

## 2 Anglerfish (Lophius budegassa, Lophius piscatorius) in subareas 4 and 6 and in Division 3.a (North Sea, Rockall and West of Scotland, Skagerrak and Kattegat)

## Assessment in 2019

The last benchmark for this stock was carried out in February 2018 (ICES, 2018). The assessment carried out at the WG follows the agreed procedure for category 3.2.0 of ICES RGLIFE datalimited stock (DLS) methods as set out in the stock annex.

ICES advice applicable to 2018 and 2019

ICES advice for 2018
ICES advise that when the precautionary approach is applied, catches in 2018 should be no more than 26408 tonnes. If discard rates do not change from the average of the last three years (20142016), this implies landings of no more than 25563 tonnes.

## ICES advice for 2019

ICES advise that when the precautionary approach is applied, catches in 2019 should be no more than 31690 tonnes.

### 2.1 General

## Stock description and management units

The anglerfish stock on the Northern Shelf is considered to occur in Division 3.a (Skagerrak and Kattegat), Subarea 4 (the North Sea) and Subarea 6 (West of Scotland plus Rockall). Anglerfish in the North Sea and Skagerrak/Kattegat were considered by this Working Group for the first time in 1999. In 2004, the WGNSDS considered the stock structure of anglerfish on a wider European scale, and found no conclusive evidence to indicate an extension of the stock area northwards to include Division 2.a. In 2013, Division 2.a was removed from WGCSE ToR.

Management applicable to 2018 and 2019

| Species: | Anglerfish | Zone: | Union waters of Ila and IV |
| :---: | :---: | :---: | :---: |
|  | Lophiidae |  | (ANF/2AC4-C) |
| Belgium |  | $573{ }^{(1)}$ |  |
| Denmark |  | $1264{ }^{1}$ ) |  |
| Germany |  | $618{ }^{(1)}$ |  |
| France |  | $118{ }^{(1)}$ |  |
| The Netherlands |  | $434\left({ }^{1}\right)$ |  |
| Sweden |  | $\left.15{ }^{1}\right)$ |  |
| United Kingdom |  | $13203\left({ }^{1}\right)$ |  |
| Union |  | $16225{ }^{(1)}$ |  |
| TAC |  | 16225 | Analytical TAC |

[^1]| Species | Anglerfish | Zone: | Norwegian waters of 4 |
| :---: | :---: | :---: | :---: |
|  | Lophiidae |  | (ANF/O4-N.) |
| Belgium |  | 51 |  |
| Denmark |  | 1305 |  |
| Germany |  | 21 |  |
| The Netherlands |  | 18 | Analytical TAC |
| United Kingdom |  | 305 | Article 3 of Regulation (EC) |
| Union |  | 1700 | No 847/96 shall not apply |
| TAC | Not relevant | Article 4 of Regulation (EC) |  |
|  |  |  | No 847/96 shall not apply |
| Species | Anglerfish | Zone: | 6; Union and international waters of |
|  | Lophiidae |  | 5.b; international waters of 7 and 14 |
|  |  |  | (ANF/56-14) |
| Belgium |  | 330 |  |
| Germany |  | 377 |  |
| Spain |  | 353 |  |
| France |  | 4059 |  |
| Ireland |  | 918 |  |
| The Netherlands |  | 318 |  |
| United Kingdom |  | 2825 |  |
| Union |  | 9180 |  |
| TAC |  | 9180 | Precautionary TAC |

COUNCIL REGULATION (EU) No 120/2018 of 23 January 2018 fixing for 2018 the fishing opportunities available to EU vessels for certain fish stocks and groups of fish stocks which are not subject to international negotiations or agreements.

| Species: | Anglerfish | Zone: | Union waters of 2.a and 4 |
| :---: | :---: | :---: | :---: |
|  | Lophiidae |  | (ANF/2AC4-C) |
| Belgium |  | $715{ }^{(1)}$ |  |
| Denmark |  | $1577{ }^{(1)}$ |  |
| Germany |  | $770{ }^{(1)}$ |  |
| France |  | $147{ }^{(1)}$ |  |
| The Netherlands |  | $541{ }^{1}$ ) |  |
| Sweden |  | $18{ }^{(1)}$ |  |
| United Kingdom |  | $16469\left(1^{1}\right)$ |  |
| Union |  | $20237{ }^{1}$ ) |  |
| TAC |  | 20237 | Analytical TAC |

${ }^{(1)}$ Special condition: of which up to $10 \%$ may be fished in: 6; Union and international waters of 5b; international waters of 12 and 14 (ANF/*56-14).

| Species | Anglerfish | Zone: | Norwegian waters of 4 |
| :---: | :---: | :---: | :---: |
|  | Lophiidae |  | (ANF/04-N.) |
| Belgium |  | 51 |  |
| Denmark |  | 1305 |  |
| Germany |  | 21 |  |
| The Netherlands |  | 18 | Analytical TAC |
| United Kingdom |  | 305 | Article 3 of Regulation (EC) |
| Union |  | 1700 | No 847/96 shall not apply |
| TAC | Not relevant |  | Article 4 of Regulation (EC) |
|  |  |  | No 847/96 shall not apply |
| Species | Anglerfish | Zone | 6; Union and international waters of 5.b; international waters of 7 and 14 |
|  | Lophiidae |  | (ANF/56-14) |
| Belgium |  | $411{ }^{1}{ }^{1}$ |  |
| Germany |  | $470{ }^{(1)}$ |  |
| Spain |  | 440 |  |
| France |  | 5067 (1) |  |
| Ireland |  | 1145 |  |
| The Netherlands |  | $396{ }^{(1)}$ |  |
| United Kingdom |  | $3524{ }^{(1)}$ |  |
| Union |  | 11453 |  |
| TAC |  | 11453 | Precautionary TAC |

${ }^{1)}$ Special condition: of which up to $5 \%$ may be fished in: Union waters of 2 a and 4 (ANF/*2AC4C).

COUNCIL REGULATION (EU) No 124/2019 of 30 January 2019 fixing for 2019 the fishing opportunities available to EU vessels for certain fish stocks and groups of fish stocks which are not subject to international negotiations or agreements.

Management of Northern Shelf anglerfish is based on separate TACs for the North Sea Subarea 4 and West of Scotland Subarea 6. There is no TAC for Skagerrak and Kattegat Division 3.a. Table 4.1 summarises the ICES advice and actual management applicable for Northern Shelf anglerfish during 2003-2019.

## Fishery description

A more detailed description of the fisheries can be found in the Stock Annex. The official national landings as reported to ICES are given in Table 4.2 and the breakdown by country in Tables 4.34.5. Minor revisions were made to tables in 2019 with updates from the ICES official and historical nominal catch statistics and the addition of the preliminary catch statistics values for 2018. Total officially reported landings of anglerfish from the Northern Shelf are shown in Figure 4.1.

## The fishery in 2018

Official landings in 2018 for subareas 6 and 4 were 20436 t ( 6211 t and 14225 t ), giving a 20\% undershoot of the combined TAC of 25405 t ( $68 \%$ and $79 \%$ TAC uptake respectively). In Subarea 6 Belgium ( $0 \%$ ), the Netherlands ( $0 \%$ ) and France ( $32 \%$ ) had noticeably low uptakes. Belgium and the Netherlands were also observed to significantly undertake their quota in Subarea $4(25 \%$ and $41 \%$ ). Denmark ( $86 \%$ ), Germany ( $84 \%$ ), and the United Kingdom all ( $86 \%$ ) decreased their Subarea 4 uptakes in comparison to 2017. The UK exceeded its quota in Subarea 6 (by 20\%) although this was a reduction from $47 \%$ in 2017. Over quota landings by individual states are most likely due to countries obtaining additional quota from other EU member states, or carrying forward unutilised quota from 2017 and using a flexibility allowance whereby $10 \%$ of Subarea 4 TAC can be utilised to reattribute landings from Subarea 6.

Uptake of EC quota in 2018, based on the preliminary officially reported landings, was as follows:

|  | TAC 6 | Lan-dings 6 | Uptake (\%) | TAC 4 <br> (Norwegian) | $\begin{aligned} & \text { TAC 2.a } \\ & \& 4 \end{aligned}$ | TAC 2.a \& 4(total) | Landings $4$ | Uptake (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 330 | - | 0\% | 51 | 573 | 624 | 156 | 25\% |
| Denmark | - | - | - | 1305 | 1264 | 2569 | 2201 | 86\% |
| France | 4059 | 1287 | 32\% | - | 118 | 118 | 142 | 120\% |
| Germany | 377 | 394 | 104\% | 21 | 618 | 639 | 536 | 84\% |
| Ireland | 918 | 878 | 96\% | - | - | - | - | - |
| Netherlands | 318 | - | 0\% | 18 | 434 | 452 | 187 | 41\% |
| Norway | - | 4 | - | - | - | - | 1267 | - |
| Russia | - | - | - | - | - | - | - | - |
| Spain | 353 | 261 | 74\% | - | - | - | - | - |
| Sweden | - | - | - | - | 15 | 15 | 25 | 167\% |
| UK (total) | 2825 | 3387 | 120\% | 305 | 13203 | 13508 | 9711 | 72\% |
| Total Union TAC | 9180 | 6207 | 68\% | 1700 | 16225 | 17925 | 12958 | 72\% |

${ }^{1}$ TAC applies to $6,5 . b$ (EC), and international waters of 7 and 14.
${ }^{2}$ Norwegian waters.

Based on data submitted to ICES, the fishery was principally prosecuted by vessels using demersal trawls (Table 4.6), targeting either white fish ( $70 \%$ of total landings by weight) or Nephrops ( $4 \%$ ). Alongside these fleets there was also a significant gillnet fishery ( $17 \%$ ), as well as an assortment of other gears in which small quantities of anglerfish are caught as bycatch. The latter have been grouped here as miscellaneous gears (9\%). Gillnets and miscellaneous gears accounted for more of the proportional split of landings across gear types in 2018 in comparison to 2017.

UK (Scottish) vessels accounted for the majority of reported anglerfish landings from the combined Northern Shelf area, taking approximately $61 \%$ of the landings overall. Scottish, Danish and Norwegian vessels took $68 \%, 15 \%$ and $9 \%$, respectively, of the North Sea (divisions 4.a-4.c) landings. Scottish, French and Irish vessels took $55 \%$, $21 \%$ and $14 \%$, respectively, of the West Coast (Subarea 6) landings. In 2013, landings were at their lowest level since the late 1980s, well below the TAC, since then they have increased by over $70 \%$. Anecdotal information on the fishery in 2018 from industry representatives is that there were fewer monks on the grounds with whitefish vessels more frequently targeting squid and some Nephrops vessels fishing central and southern North Sea grounds with lower abundances of monkfish. Port officials noted that there was a larger grade of monks being landed, as well as fewer small monks 'frogs' landed despite the low quota uptake. This is a stark change to the industry feedback of 2016-2017 where the high abundance of fish on the ground and the high-value sizeable individuals of the 2014 cohort were resulting in fishermen becoming more selective during on-board processing with suspected increased discarding and high-grading.

Landings in Division 3.a are not regulated: Table 4.5 shows the official landings which had fluctuated between 400-500 t from 2005-2015, between 2016 and 2018 they have increased significantly, on average $26 \%$ per year with the 2018 landings now at 914 t .

### 2.2 Data

## Landings

National landings data as reported to ICES and Working Group estimates of total landings are given in Table 4.2. The working group procedures used to determine the total international landings numbers and weights-at-length are documented in the stock annex. It is acknowledged that throughout the landings time-series, there have consistently been differences between the total official landings and the landings as estimated by the WG. This is likely due to differences in the data provided to the WG by national scientists and administrators.

Due to restrictive TACs, the likelihood of misreporting and underreporting of anglerfish landings in the past is considered to have been high, particularly during the period 2003-2005. During the benchmark at WKROUND (ICES, 2013), it was agreed that recent landings are likely to be more accurate from 2006 due to, i) less restrictive TACs, ii) the introduction of buyers and sellers legislation in the UK and Ireland and iii) the offshore gillnet fishery for anglerfish historically conducted by Spanish flagged vessels and thought to under-report landings, being much reduced. Anecdotal reports from fisheries offices and catch sampling staff suggest that towards the end of 2016 and into 2017 the high abundance of anglerfish on the grounds, and the restrictive quota were leading to an increase in suspected misreporting, discarding and black landings. There was no new information in 2019 to suggest that these suspected practices continued into 2018, and the substantially lower quota uptake of 2018 may indicate that the incentives for this behaviour are no longer prevalent. During the period 2005-2010, landings data were not provided to the Working Group by some of the major nations exploiting the fishery; however, the recent data call for the WKANGLER benchmark (2018) has meant that WG estimates of Subareas 6 and 4 and Division 3.a landings have now been calculated for this period.

## Discards

Prior to the recent WKANGLER benchmark (2018) discard estimates have only been available within InterCatch since 2012. Following the WKANGLER data call discard information is now available for some fleets since 2002; however, discard information from UK (Scotland) is not available before 2007. The discard estimates that are available from other nations for the 20022006 period are substantially higher than the later UK (Scottish) rates. Given that these fleets represent proportionally less of the landings the discards pre-2007 are considered to be nonrepresentative of the overall fishery (WKANGLER 2018). The breakdown of landings and discards by main gear group and area for 2018 is given in Table 4.6. Discard data indicate that discarding in this fishery is relatively low due to high market value and no MLS. Overall discarding was $1.5 \%$ of total catch in 2018, a reduction on the 2017 rate of $3.4 \%$. Demersal TR2 trawlers had the highest discard rate due to more restrictive quota share, $9.9 \%$ in 2018 down from $20.9 \%$ in 2017 and 43.9 in 2016. In comparison TR1 trawlers, gillnets and miscellaneous gear types had much lower rates of $1.0 \%, 1.7 \%$ and $1.5 \%$ respectively. Discards in Subarea 4 ( 113 t ) were lower than in Subarea $6(170 \mathrm{t})$, by weight and proportion ( $0.8 \%$ and $2.6 \%$ ).

Figures $4.3(\mathrm{a}-\mathrm{c})$ show the percentage of landed weight by fleet, country and area. Length-frequency samples for catch in 2018 were submitted by Belgium, Denmark, France, Germany, Ireland, Norway, UK (England \& Wales) and the UK (Scotland). There was good coverage of both the demersal TR1 and TR2 fleets in Subarea 4 and Division 6.a. However once again there were poor levels of sampling for the TR1 fleet in Division 27.6.b with only five samples for landings (Ireland: 307 fish) and 14 for discards (UK (Scotland): four samples, 167 fish and Ireland: 12 samples, 316 fish). Denmark and Norway provided samples for gillnet fleets in Subarea 4 and Division 3.a. There were no samples from UK-flag gillnet vessels which alone accounted for approximately $8 \%$ of all landings.

Discard data were used in the provision of catch advice until 2018 in line with the DLS approach (ICES, 2012). From 1 January 2019, all caught anglerfish must be retained under the discard ban of the EU landing obligation regulation.

## Biological

An anglerfish ageing exchange was held in 2011 to investigate the possibility of the collation of an international landings-at-age dataset of hard structure age readings, however little agreement was found between methods or readers. This was acknowledged in the findings of the WKROUND report on current assessment and issues with data and assessment of this stock (ICES, 2013). Further to this, discussions at WKANGLER established that few countries are actively reading anglerfish hard structures, although they continue to be collected, processed and stored. It is unlikely that any developments in regards to an agreed reading criterion will be made in the near future.

## Research vessel surveys

The 2019 SIAMISS-Q2 survey is described in detail in the Stock Annex and the most recent results of the 2019 SIAMISS-Q2 can be found in WDXX of the 2019 WGCSE report. This is a targeted anglerfish survey using commercial gear, covering subareas 4 and 6 . The abundance and biomass estimates from the surveys are presented in Tables 4.7 and 4.8. The total biomass estimates for the Northern Shelf in 2018 and 2019 were 77661 and 58575 t respectively.

Total numbers of anglerfish increased in the years 2011-2015; however, from 2016 onwards there has been a fluctuating but declining trend with survey division estimates becoming more divergent (Table 4.8 and Figure 4.6). The most recent year (2019) has seen an increase in numbers driven by increases in both Subarea 4 and Division 6.a. Total biomass of anglerfish increased from 2011-2017 after which there has been two years of decline. In 2019, all surveyed divisions saw a decrease in biomass. The substantial increase in numbers (2014-2015) and biomass (20142017) was due to a large number of small fish having entered the stock in 2013 (Figure 4.6). The scale of this year class has not previously been seen in the SIAMISS-Q2 survey (for years for which length data are currently available 2007-2018) (Figure 4.8). Whilst this year class was clearly identifiable in 2014 and 2015, in the total survey abundance-at-length (Figure 4.8) 2016 was the first year in which the year class's contribution to total biomass-at-length is clearly observed (Figure 5.2.9). This cohort mode is less apparent in the 2019 weight-at-length (Figure 5.2.9) with the peak of the distribution now more plateaued between 65-85 cm.

After a period of low surveyed abundance in both Subarea 4 and Division 6.a for the years 20092012, there has been a significant increasing trend in the years following, however 2016 saw the first decline in abundance in five years. Whilst the abundance and biomass of anglerfish in Subarea 4 and Division 6.a have tracked each other relatively well over the time-series, in 2015-2016 the areas have shown divergent trends with a decline in $6 . a$ and 4 continuing to increase. Both numbers and biomass estimates for Subarea 4 have fallen from 2018, two years after the same trend was observed in Division 6.a suggesting a possible time lag of decline. Although numbers of fish in Division 6.b have remained relatively stable since 2012 when all three surveyed areas had similar abundance estimates, the biomass in Division $6 . b$ has increased steadily with the exception of a decline in 2017. The 2018 estimate of biomass in Division $6 . \mathrm{b}$ was more than double the time-series average of 11208 kt; however the much lower estimate in 2019 is now more in line with the previous trajectory of increase.

Estimates of the ratio of survey biomass between subareas 4 and 6 have fluctuated around 1:1, (time-series average of $47 \%$ in Subarea 4, Table 4.7). The proportion of biomass in Subarea 4 had been steadily increasing since 2013; however, 2017 saw a slight decrease followed by a marked decline in 2018 to a time-series low of $37 \%$ (Figure 4.10). 2019 has seen an increase in the proportion of biomass in Subarea 4 moving back towards a 1:1 split.

Additional survey indexes were developed during the WKANGLER 2018 benchmark after revisiting the anglerfish abundance of several surveys within the stock area (ICES, 2018). Mean weight per hour for both the SCW-IBTS Q1 and Q4 surveys has declined in 2018 following recent time-series highs in 2017 and 2016 respectively (Figure 4.13) which reflects the SIAMISS-Q2 biomass trend (Figure 4.6). The ROCKALL index (Figure 4.13) shows an increasing trend since 2005 with a significant peak in 2012 followed by a short period of decline before a continuation of the increasing trajectory from 2016 to present. This increase in biomass is attributable to large fish $>60 \mathrm{~cm}$. Although the SIAMISS-Q2 biomass time-series for Division 6.b shows less year to year fluctuation than the Rockall index, the increasing trend and magnitude of change for the 20052018 period are very similar. In Subarea 4, the NS-IBTS-Q1 and Q3 indexes show declining mean weights per hour for the recent 3-4 years across all length groupings (Figure 4.14). This contradicts the SIAMISS-Q2 biomass series, which continued to increase until 2017 before a marked decline in 2018.

## Commercial catch-effort data

Trends in nominal international fishing effort in the North Sea and Eastern Channel and the West of Scotland collated by STECF for the Evaluation of Fishing Effort Regimes in European Waters are shown in Figure 4.2. Since 2014, there have been slight increases in TR effort in both the North Sea and West of Scotland, with effort across all gears in the North Sea stable or reducing since

2012 and in the West of Scotland increasing in the past two years driven by marked increases in trawl fisheries. Data for 2017 has not yet been released by STECF although a significant change in the overall observed trend of anglerfish fleets is not anticipated with the introduction of 2017 data.

There is now a time-series of commercial catch-at-length data for 2002-2018 (Figure 4.4). The spread of lengths in the landings distributions are wider during the period 2012-2014 after which the distributions are steeper with singular peaks. In 2015, the strong 2013 cohort entered the fishery producing a markedly different catch composition of lengths with the bulk of landings being between 30 and 50 cm in length with steep tails either side. Discard rates are lower from 2015 onwards however, the landings of $<30 \mathrm{~cm}$ fish were also lower, suggesting this reduction could be a combination of catch composition and the increase in quota availability. The distribution of lengths in the landings in 2018 has a wider spread to its peak than in recent years likely due to contribution of the 2014 cohort, which are now larger individuals.

### 2.3 Historical stock development

There has been no analytic assessment of Northern Shelf anglerfish since 2003, due to a combination of unreliable commercial data, landings misreporting, uncertain effort data and poor catchability of anglerfish in traditional research surveys. The Scottish Irish anglerfish and megrim industry science survey (SIAMISS-Q2) initiated by Marine Scotland Science in 2005, along with official logbook data and tally-book data schemes have addressed some of these issues, providing valuable information to fishery managers as well as minimum absolute abundance and biomass estimates annually. Since 2012 assessment has followed the ICES RGLIFE data-limited stock (DLS) 3.2.0 method of survey based indicative trends (ICES, 2012).

### 2.4 Short-term projections

In the absence of an age-based assessment, there are no short-term projections for this stock.

### 2.5 Biological reference points

Precautionary approach reference points.

|  | Type | Value | Technical basis |
| :---: | :---: | :---: | :---: |
| Precautionary approach | $\mathrm{Blim}^{\text {l }}$ | Not defined | There is currently no biological basis for defining $\mathrm{Bl}_{\text {lim }}$ |
|  | $\mathrm{B}_{\mathrm{pa}}$ | Not defined |  |
|  | $\mathrm{F}_{\text {lim }}$ | Not defined | There is currently no biological basis for defining $\mathrm{F}_{\text {lim }}$ |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.30 | $\mathrm{F}_{35 \% \text { SPR }}=0.30$. This fishing mortality corresponds to $35 \%$ of the unfished SSB/R. It is considered to be an approximation of $\mathrm{F}_{\text {MSY }}$. |
| Targets | $\mathrm{F}_{\mathrm{y}}$ | Not defined |  |

(unchanged since 1998).

## Yield-per-recruit analysis and harvest rates

One suggested method for future assessment is a Nephrops-like harvest-ratio approach, which creates a catch-options table based on a range of harvest ratios. However to date no MSY reference points have been determined for Northern shelf anglerfish despite further exploration (Holah, H., 2017). Limited data, dome-shaped selectivity and uncertain life-history parameters continue to be inhibiting factors. Previous attempts to determine suitable harvesting rates, based on a yield-per-recruit analysis, estimated $\mathrm{F}_{\text {max }}$ to be 0.19 (ICES, 2004). The southern stock has recently been benchmarked and an $\mathrm{F}_{\text {MAX }}$ of 0.28 was adopted (ICES, 2012b). This needs to be revisited for this stock. In the case of Nephrops the technical basis for MSY Btrigger is the bias-adjusted lowest observed UWTV survey estimate of abundance, however for anglerfish, whilst abundances from SIAMISS-Q2 were initially intended to be an absolute measure of abundance they are now considered to be only a relative index so this may not be appropriate.

Figure 4.11 shows mean standardised harvest rate by both weight and number of individuals, whilst there are no reference levels to relate these harvest rates to, trends can still be interpreted. In 2018, both the harvest rate of biomass and harvest rate as number of individuals have increased. The marked fall in harvest rate by number from 2013-2014 is likely due to the influx of the substantial 2013 year class and not a change in fishing behaviour. As a result of the 2013 year classes now reaching exploitable length, the harvest rate in 2016 has increased. It may be more appropriate to use a harvest rate which is measured over a given length range of commercially exploitable fish.

### 2.6 Management plans

There is no management plan for this stock.

### 2.7 Uncertainties and bias in assessment and forecast

The WGCSE has previously attempted assessments of the anglerfish stock(s) within its remit using a number of different approaches. As yet none have proved entirely satisfactory. The catch-at-length analysis used in previous years appears to have addressed a number of the suspected problems with the data due to the rapid development of the fishery, and has also provided a satisfactory fit to the catch-at-length distribution data. However, since 2003, the WG has been unable to present an analytic assessment due to the lack of reliable fishery and insufficient survey information, and in addition, it is not known to what extent the dynamic pool assumptions of the traditional assessment model are valid for anglerfish. A catch-at-age model was presented to two benchmark working groups (WKFLAT 2012 and WKROUND 2013) but was not accepted due to concerns over age reading. The SPiCT and ASPIC surplus production models were explored at the WKANGLER benchmark (2018) and whilst the models converged, the models were unstable and the uncertainty was large. This is most likely due to the lack of contrast in the catch data.

## Commercial data

For a number of years the WG has expressed concerns over the quality of the commercial catch-at-length data because of:

- Accuracy of landings statistics due to species and area misreporting (historically an issue between 1998-2005 and anecdotally again in 2016).
- Lack of information on total catch and catch composition of gillnetters operating on the continental slope to the northwest of the British Isles (See the stock annex for further details of this fishery).


## Survey data

There are still several factors which make the survey estimates likely to be underestimates or minimum estimates. Firstly, although experiments have been carried out to estimate escapes from under the footrope, and a model applied to account for this component of catchability, the estimates of smaller anglerfish still look to be underestimated (Figure 4.7). This could be due to either a net selectivity issue, or an availability [to the trawl] issue, as it is known that younger fish occur in shallower water (Hislop et al., 2001), or both. Secondly, the area considered is not complete, as the survey does not cover some of Division 4.a and none of 4.b or 4.c. However, numbers are thought to be low in these areas. A comparison of mean length in the commercial catch (Figure 4.12) and in the SIAMISS-Q2 and NS-IBTS surveys suggests that the selectivity of the commercial fleet and the angler SIAMISS-Q2 survey gear are similar (before the survey estimation procedure of corrections is applied).

## Biological information

Knowledge of the biology of anglerfish has improved, with some basic biological parameters suitable for use in future assessments, such as mean weight-at-length in the stock, now available from the industry-science survey data. Difficulties still remain in finding mature females. A further discussion of the biology can be found in the stock annex.

Life-history parameters of the anglerfish species Lophius piscatorius and Lophius Budegassa in the Northeast Atlantic were reviewed at the WKANGLER benchmark (2018) with appropriate ranges of natural mortality $(\mathrm{M})$ discussed and new approaches to estimating age from mixture modelling of length distributions presented (see WKANGLER 2018 report for further details).

## Stock structure

Currently, anglerfish on the Northern Shelf are split into Subarea 6 (including 5.b (EC), 7 and 14) and the North Sea (and 2.a (EC)) for management purposes. However, genetic studies have found no evidence of separate stocks over these two regions (including Rockall) and particletracking studies have indicated interchange of larvae between the two areas (Hislop et al., 2001). So at previous WGs, assessments have been made for the whole Northern Shelf area combined. In fact, both microsatellite DNA analysis (O'Sullivan et al., 2005) and particle tracking studies carried out as part of EC 98/096 (Anon, 2001) also suggested that anglerfish from further south (Subarea 7) could also be part of the same stock.

At present, the stock is assessed for the two anglerfish species L. piscatorius and L. budegassa combined, despite differing life-history characteristics and overlap in spatial distribution. This has been the case due to the black anglerfish (L. Budegassa) proportionally representing only around $10 \%$ of the estimated stock biomass from the SIAMISS-Q2 survey and that the Scottish fleet land the two species for sale combined as "monkfish". Given that the proportion of black anglerfish has been as high as $28 \%$ in Division 6.a and that the Scottish market sampling programme records to species level, a splitting out of black anglerfish in this stock may be a consideration for a future benchmark.

### 2.8 Recommendations for next Benchmark

This stock was last benchmarked in February 2018 at WKANGLER. The recommendations to be carried forward following WKANGLER are the following tasks:

- Investigate length-based stock assessment using, for example, the SS3 approach applied to southern anglerfish stocks.
- Investigate growth models appropriate for anglerfish subareas 4 and 6.
- Investigate an age-aggregated production/depletion model.
- Determine the best way to incorporate Lophius budegassa into assessment and advice.

The WKANGLER data call led to the compilation of commercial sampling data (length, age, weight) previously held internationally, to construct a historical catch-at-length dataset for 2002 to present. At this stage, the focus is currently to utilise this more complete dataset to develop a suitable assessment model for this stock.

### 2.9 Management considerations

Up to and including 2011, ICES provided qualitative advice regarding the future exploitation of 'data-limited' stocks where there was either limited knowledge of their biology or a lack of data on their exploitation. However, in response to a strong interest from advice recipients to base advice on the information available, ICES developed the data-limited stocks (DLS) approach framework, for which anglerfish is a category 3 data-limited stock. This requires considering the application of an uncertainty cap and/or precautionary buffer to a survey adjusted status quo catch at each annual advice draft.

A comparison of mean biomass estimates from the SIAMISS-Q2 surveys (Table 4.9) shows that the mean biomass in subareas 4 and 6 combined has increased by $25.0 \%$ from 2014-2016 to 20172018. Application of the uncertainty cap implied advice for catches in 2019 to be no more than $20 \%$ greater than the previously advised catch. The stability observed in international effort timeseries by the main fisheries in the stock area since 2003 meant that a precautionary buffer should not be applied.

The TACs in subareas 4 (including Norwegian waters) and 6 until 2010 were split $67: 33 \%$, since 2011 they have been split $64: 36 \%$. In $2018,10 \%$ of the TAC for 4 and 2.a could be taken from Division 5.b, or subareas 6,7 and 9 . Over the survey time-series, the stock has been fairly evenly distributed between 4:6, the split has fluctuated around 50:50 (47\% on average) (Table 4.7 and Figure 4.10) however in 2018 there was a significant decrease to $38 \%$ increasing to $40 \%$ in 2019. Note that the North Sea is only partially surveyed: however, the area covered does encompass most of the distribution of anglerfish.

Ideally, the management of the fishery should be based on a specific plan, or harvest control rule, after an evaluation of various stakeholder-led suggestions of alternative options. This still needs to be pursued in consultation with stakeholders such as the North Western Waters Advisory Council.

### 2.10 References

Anon. 2001. The distribution and biology of anglerfish and megrim in waters to the west of Scotland. EC Study Contract 98/096 Final Report August 2001.

Clarke, E., Buch, T., Gillsepie-Mules, R., Drewery, J., Kelly, E and Gerritsen, H. 2018. WD for WGCSE October 2018: SIAMISS Estimates of Anglerfish Biomass in subareas 4 and 6 for 2018. Working document to the Celtic Seas Ecosystem Working Group, ICES 2017, 14 pp.
COUNCIL REGULATION (EU). 2018. No 2018/120 of 23 January 2018 fixing for 2018 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union vessels, in certain non-Union waters, amending Regulation (EU) No 43/2014 and repealing Regulation (EU) No 2017/127.

COUNCIL REGULATION (EU). 2019. No 124/2019 of 30 January 2019 fixing for 2019 the fishing opportunities available to EU vessels for certain fish stocks and groups of fish stocks which are not subject to international negotiations or agreements.
Hislop, J. R. G., Gallego, A., Heath, M. R., Kennedy, F. M., Reeves, S. A., and Wright, P. J. 2001. A synthesis of the early life history of the anglerfish, Lophius piscatorius (Linnaeus, 1758) in Northern British waters. ICES Journal of Marine Science 58:70-86.

Historical Nominal Catches 1950-2010. Version 30-11-2011. Accessed 13-05-2019 via http://ices.dk/marine-data/dataset-collections/Pages/Fish-catch-and-stock-assessment.aspx. ICES, Copenhagen.
Holah, H. 2017. Length based indicators and SPiCT in relation to reference points for Anglerfish in subareas 4 and 6, and in Division 3.a (North Sea, Rockall and West of Scotland, Skagerrak and Kattegat) (anf.27.3a46). Working document. Celtic Seas Ecosystem Working Group, ICES 2017.

ICES. 2004. Report of the Working Group on the Assessment of Northern Shelf Demersal Stocks. ICES CM 2004/ACFM:01. 558 pp.
ICES. 2012. Report of the ICES Advisory Committee on ICES implementation of Advice for Data-limited Stocks in 2012 in its 2012 Advice. ICES CM 2012/ACOM:68. 42 pp.

ICES. 2012b. Report of the Benchmark Workshop on Flatfish Species and Anglerfish (WKFLAT), 1-8 March 2012, Bilbao, Spain. ICES CM 2012/ACOM:46. 283 pp.

ICES. 2013. Report of the Benchmark Workshop on Roundfish Stocks (WKROUND), 4-8 February, Aberdeen. ICES CM 2013 / ACOM:47 213 pp.
ICES. 2016. Report of the Workshop to consider MSY proxies for stocks in ICES category 3 and 4 stocks in Western Waters (WKProxy), 3-6 November 2015, ICES Headquarters, Copenhagen. ICES CM 2015/ACOM:61. 183 pp.

ICES. 2018. Report of the Benchmark Workshop on Anglerfish Stocks in the ICES Area (WKANGLER), 1216 February 2018, Copenhagen, Denmark. ICES CM 2018/ACOM:31. 177 pp.
Official Nominal Catches 2006-2014. Version 08-05-2019. Accessed 13-05-2019 via http://ices.dk/marine-data/dataset-collections/Pages/Fish-catch-and-stock-assessment.aspx. ICES, Copenhagen.

O'Sullivan, M., Wright, P., Vespoor, E., Knox, D. and Piertny, S. 2005. Microsatellite DNA polymorphism indicates an absence of population structure in monkfish Lophius piscatorius in its northern distribution. ICES CM2005/T:18 (poster).
Scientific, Technical and Economic Committee for Fisheries (STECF). 2017 - Fisheries Dependent - Information - Classic (STECF-17-09). Publications Office of the European Union, Luxembourg 2017, ISBN 978-92-79-67481-5, doi:10.2760/561459, JRC107598.

Table 4.1. ICES advice and actual management applicable for Northern Shelf anglerfish for 2003 onwards.

| YEAR | SINGLE STOCK EX- <br> PLOITATION <br> BOUNDARY | BASIS | WEST OF SCOTLAND (6.a-6.b) | NORTH SEA (4.a-4.c) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

All values raised to nearest tonne.
${ }^{1)}$ Advice for Division 3.a, Subarea 4 and Subarea 6.a combined.
${ }^{2}$ ) Advice for Division 3.a, Subarea 4 and Subarea 6 combined.
${ }^{3}$ ) Advice for Division 2.a, Division 3.a, Subarea 4 and Subarea 6 combined.
${ }^{4)}$ TAC applies to $5 . b(E C), 6,7$ and 14.
5) TAC applies to $2 . a \& 4$ (EC).
${ }^{6)}$ of which up to $10 \%$ may be fished in: 5.b(EC), 6, 7 and 14 .
${ }^{(7)}$ Landings including raised discards.

Although there is no minimum landing size for this species, there is an EU minimum weight of 500 g for marketing purposes (EC Regulation 2406/96).
An additional quota of 1500 t was also available for EU vessels fishing in the Norwegian zone of Subarea 4 in 2011-2018 which was increased to 1700 t in 2018.

Table 4.2. Anglerfish on the Northern Shelf (3.a, $4 \& 6$ ). Total official landings by area (tonnes).

|  | 3.a | 4.a | 4.b | 4.c | 6.a | 6.b | 4 | 6 | Total $(3 . A, 4,6)$ | WG Landings | WG Discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1973 | 140 | 2085 | 575 | 41 | 9221 | 127 | 2701 | 9348 | 12189 | - | - |
| 1974 | 202 | 2737 | 1171 | 39 | 3217 | 435 | 3947 | 3652 | 7801 | - | - |
| 1975 | 291 | 2887 | 1864 | 59 | 3122 | 76 | 4810 | 3198 | 8299 | - | - |
| 1976 | 641 | 3624 | 1252 | 49 | 3383 | 72 | 4925 | 3455 | 9021 | - | - |
| 1977 | 643 | 3264 | 1278 | 54 | 3457 | 78 | 4596 | 3535 | 8774 | - | - |
| 1978 | 509 | 3111 | 1260 | 72 | 3117 | 103 | 4443 | 3220 | 8172 | - | - |
| 1979 | 687 | 2972 | 1578 | 112 | 2745 | 29 | 4662 | 2774 | 8123 | - | - |
| 1980 | 652 | 3450 | 1374 | 175 | 2634 | 200 | 4999 | 2834 | 8485 | - | - |
| 1981 | 549 | 2472 | 752 | 132 | 1387 | 331 | 3356 | 1718 | 5623 | - | - |
| 1982 | 529 | 2214 | 654 | 99 | 3154 | 454 | 2967 | 3608 | 7104 | - | - |
| 1983 | 506 | 2465 | 1540 | 181 | 3417 | 433 | 4186 | 3850 | 8542 | - | - |
| 1984 | 568 | 3874 | 1803 | 188 | 3935 | 707 | 5865 | 4642 | 11075 | - | - |
| 1985 | 578 | 4569 | 1798 | 77 | 4043 | 1013 | 6444 | 5056 | 12078 | - | - |
| 1986 | 524 | 5594 | 1762 | 47 | 3090 | 1326 | 7403 | 4416 | 12343 | - | - |
| 1987 | 589 | 7705 | 1768 | 66 | 3955 | 1294 | 9539 | 5249 | 15377 | - | - |
| 1988 | 347 | 7737 | 2061 | 95 | 6003 | 1730 | 9893 | 7733 | 17973 | - | - |
| 1989 | 334 | 7868 | 2121 | 86 | 5729 | 313 | 10075 | 6042 | 16451 | - | - |
| 1990 | 570 | 8387 | 2177 | 34 | 5615 | 822 | 10598 | 6437 | 17605 | - | - |
| 1991 | 595 | 9235 | 2522 | 26 | 5061 | 923 | 11790 | 5984 | 18369 | 17441 | - |
| 1992 | 938 | 10209 | 3053 | 39 | 5479 | 1089 | 13301 | 6568 | 20807 | 21872 | - |
| 1993 | 843 | 12309 | 3143 | 66 | 5553 | 681 | 15519 | 6234 | 22596 | 23971 | - |
| 1994 | 811 | 14505 | 3445 | 210 | 5273 | 909 | 18162 | 6182 | 25155 | 25057 | - |
| 1995 | 823 | 17891 | 2627 | 402 | 6354 | 958 | 20920 | 7312 | 29055 | 28913 | - |
| 1996 | 702 | 25176 | 1847 | 304 | 6408 | 602 | 27327 | 7010 | 35039 | 35100 | - |
| 1997 | 776 | 23425 | 2172 | 160 | 5330 | 990 | 25757 | 6320 | 32853 | 32728 | - |
| 1998 | 626 | 16859 | 2088 | 78 | 4506 | 1313 | 19026 | 5819 | 25471 | 25293 | - |
| 1999 | 660 | 13344 | 1517 | 24 | 4284 | 1401 | 14885 | 5685 | 21230 | 21854 | - |
| 2000 | 602 | 12338 | 1617 | 31 | 3311 | 1074 | 13986 | 4385 | 18973 | 19682 | - |

$\left.\left.\begin{array}{llllllllllll}\hline & \mathbf{3 . a} & 4 . a & 4 . b & 4 . c & 6 . a & 6 . b & 4 & 6 & \text { Total } \\ \text { (3.A, 4,6) }\end{array}\right] \begin{array}{l}\text { WG Land- } \\ \text { ings } \\ \text { cards }\end{array}\right\}$
*Preliminary.

Table 4．3．Anglerfish in Subarea 6．Nominal landings（ $\mathbf{t}$ ）as officially reported to ICES．

## Division 6．a（West of Scotland）

|  | ন্ন | N్구 | N | す | 늑ㄱ간 | இั | 人) | がㄱㄱ | 악 | O | O-N | No | OO N | O | ONO | O O | No | Oio | Oio | $$ | $\underset{\sim}{7}$ | $\underset{\sim}{\sim}$ | $\stackrel{m}{N}$ | $\underset{\sim}{\text { İ }}$ | $\stackrel{\sim}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{N}{1} \end{aligned}$ | $\stackrel{N}{i}$ | $\stackrel{*}{*}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 3 | 2 | 9 | 6 | 5 | － | 5 | 2 | － | － | ＋ | ＋ | － | ＋ | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Denmark | 1 | 3 | 4 | 5 | 10 | 4 | 1 | 2 | 1 | ＋ | ＋ | － | ＋ | ＋ | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Faroe Is． | － | － | － | － | － | － | － | － | － | － | － | － | － | 2 | 2 | 3 | 2 | 1 | 2 | 4 | 1 | － | － | － | ＋ | 1 | － | － |
| France | 1910 | 2308 | 2467 | 2382 | 2648 | 2899 | 2058 | 1634 | 1814 | 1132 | 943 | 739 | 1212 | 1191 | 1396 | 1314 | 1764 | 1746 | 1513 | 1206 | 1168 | 1166 | 1114 | 1098 | 1107 | 1734 | 1882 | 1287 |
| Germany | 1 | 2 | 60 | 67 | 77 | 35 | 72 | 137 | 50 | 39 | 11 | 3 | 27 | 39 | 39 | 1 | － | 54 | 79 | 79 | 59 | 63 | 48 | 85 | 63 | 81 | 79 | 127 |
| Ireland | 250 | 403 | 428 | 303 | 720 | 717 | 625 | 749 | 617 | 515 | 475 | 304 | 322 | 219 | 356 | 392 | 470 | 295 | 328 | 510 | 488 | 346 | 336 | 410 | 446 | 581 | 579 | 596 |
| Nether－ | － | － | － | － | － | － | 27 | 1 | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Norway | 6 | 14 | 8 | 6 | 4 | 4 | 1 | 3 | 1 | 3 | 2 | 1 | ＋ | ＋ | 1 | 1 | 1 | 2 | ＋ | 2 | 1 | ＋ | 1 | 1 | 1 | 1 | 1 | 2 |
| Spain | 7 | 11 | 8 | 1 | 37 | 33 | 63 | 86 | 53 | 82 | 70 | 101 | 196 | 110 | 83 | 76 | 3 | 174 | 185 | 197 | 138 | 69 | 123 | 54 | 130 | 178 | 173 | 218 |
| UK（E，W\＆NI） | 270 | 351 | 223 | 370 | 320 | 201 | 156 | 119 | 60 | 44 | 40 | 32 | 31 | 30 | 20 | 24 | 42 | 5 | 12 | 3 | － | 12 | 6 | － | － | － | － | － |
| UK（Scot．） | 2613 | 2385 | 2346 | 2133 | 2533 | 2515 | 2322 | 1773 | 1688 | 1496 | 1119 | 1100 | 705 | 862 | 1127 | 974 | 1071 | 1096 | 864 | 1040 | － | 1179 | 1038 | － | － | － | － | － |
| UK（total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | 1016 | 1191 | 1044 | 962 | 1643 | 2062 | 2311 | 2061 |
| Total | 5061 | 5479 | 5553 | 5273 | 6354 | 6408 | 5330 | 4506 | 4284 | 3311 | 2660 | 2280 | 2493 | 2453 | 3024 | 2785 | 3353 | 3373 | 2984 | 3040 | 2871 | 2835 | 2667 | 2610 | 3290 | 4638 | 5024 | 4292 |
| Unallocated | 296 | 2638 | 3816 | 2766 | 5112 | 11148 | 7506 | 5234 | 3799 | 3114 | 2068 | －187 | －2 | －16 | 8 | 74 | －145 | －332 | －190 | －56 | －62 | －91 | －115 | － | －68 | 58 | －12 | －119 |
| As used by | 5357 | 8117 | 9369 | 8039 | 11466 | 17556 | 12836 | 9740 | 8083 | 6425 | 4728 | 2467 | 2495 | 2469 | 3016 | 2711 | 3498 | 3705 | 3174 | 3096 | 2933 | 2926 | 2782 | 4205 | 3358 | 4580 | 5036 | 4411 |

Table 4．3．Continued．Anglerfish in Subarea 6．Nominal landings（t）as officially reported to ICES．
Division 6．b（Rockall）／${ }^{\text {P }}$ Preliminary．

|  | ন্ন | নָ | ুু | すু | ূু | இั | ন্ত্ন | が | 욱 | Oి | O్రై | Ò | O్N | ষ্ণ | 菅 | O O | 웅 | O잉 | Oì | Oi | $\underset{\sim}{7}$ | $\underset{\sim}{\text { N}}$ | $\stackrel{\sim}{\sim}$ | $\underset{\sim}{\underset{\sim}{A}}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{0}{\circ}$ | $\stackrel{N}{N}$ | $\stackrel{*}{\stackrel{*}{\sim}} \stackrel{\stackrel{\sim}{\sim}}{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Is． | － | 2 | － | － | － | 15 | 4 | 2 | 2 | － | 1 | － | － | － | － | － | － | 1 | 4 | 8 | － | 5 | － | 1 | ＋ | ＋ | ＋ | － |
| France | － | － | 29 | － | － | － | 1 | 1 | － | 48 | 192 | 43 | 191 | 175 | 293 | 224 | 327 | 327 | 339 | 168 | 508 | 456 | 663 | 148 | 219 | － | － | － |
| Germany | － | － | 103 | 73 | 83 | 78 | 177 | 132 | 144 | 119 | 67 | 35 | 64 | 66 | 77 | 72 | 222 | 93 | 132 | 87 | 90 | 79 | 88 | 66 | 139 | 177 | 167 | 266 |
| Ireland | 272 | 417 | 96 | 135 | 133 | 90 | 139 | 130 | 75 | 81 | 134 | 51 | 26 | 13 | 35 | 53 | 70 | 76 | 91 | 107 | 108 | 235 | 237 | 162 | 156 | 160 | 214 | 282 |
| Norway | 18 | 10 | 17 | 24 | 14 | 11 | 4 | 6 | 5 | 11 | 5 | 3 | 6 | 5 | 4 | 6 | 7 | 5 | 9 | 12 | 7 | 5 | 9 | 3 | 6 | 11 | 4 | 1 |
| Portugal | － | － | － | 132 | 128 | － | 91 | 413 | 429 | 20 | 18 | 8 | 4 | 19 | 63 | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Russia | － | － | － | － | － | － | － | － | － | － | 1 | － | － | 2 | 4 | 1 | 1 | 35 | － | － | － | － | － | 1 | 2 | － | 2 | － |
| Spain | 333 | 263 | 178 | 214 | 296 | 196 | 171 | 252 | 291 | 149 | 327 | 128 | 59 | 43 | 34 | 36 | 12 | 85 | 57 | 32 | 29 | 36 | － | 27 | 119 | 56 | 118 | 43 |
| UK（E，W\＆NI） | 99 | 173 | 76 | 50 | 105 | 144 | 247 | 188 | 111 | 272 | 197 | 133 | 133 | 54 | 93 | 46 | － | 1 | 48 | 15 | － | 120 | 395 | － | － | － | － | － |
| UK（Scot） | 201 | 224 | 182 | 281 | 199 | 68 | 156 | 189 | 344 | 374 | 367 | 317 | 160 | 294 | 355 | 477 | － | 624 | 1141 | 1177 | － | 895 | 732 | － | － | － | － | － |
| UK（total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | 622 | － | － | － | 1129 | 1015 | 1127 | 1347 | 1081 | 1018 | 999 | 1326 |
| Total | 923 | 1089 | 681 | 909 | 958 | 602 | 990 | 1313 | 1401 | 1074 | 1309 | 718 | 643 | 671 | 958 | 915 | 1261 | 1246 | 1820 | 1606 | 1871 | 1831 | 2123 | 1754 | 1723 | 1423 | 1504 | 1918 |
| Unallocated | － | － | － | －132 | －128 | － | －91 | －413 | －9 | 17 | －178 | 210 | 70 | 10 | 227 | 136 | 282 | 104 | －198 | 791 | 111 | 385 | 178 | 80 | 74 | 37 | 80 | －26 |
| As used by WG | 923 | 1089 | 681 | 777 | 830 | 602 | 899 | 900 | 1392 | 1091 | 1131 | 508 | 573 | 661 | 731 | 779 | 979 | 1142 | 2018 | 815 | 1760 | 1446 | 1945 | 1674 | 1649 | 1386 | 1424 | 1944 |

Table 4．3．Continued．Anglerfish in Subarea 6．Nominal landings（t）as officially reported to ICES．
Subarea 6 （West of Scotland and Rockall）．＾indicates landings assigned to subarea 6 but not to a division．／＊Preliminary．

|  | ন্ন | ুㅜㄱ | $\stackrel{\Pi}{\sigma}$ | す | 先 | Н | 욱 | $\stackrel{\circ}{\underset{\sim}{\circ}}$ | 胢 | 으N | $\underset{\sim}{\text { OH}}$ | 우N | OiN | O | 苜 | O O N | $\underset{\sim}{\text { O}}$ | OiN | Oi | $\stackrel{\circ}{\circ}$ | $\underset{\sim}{7}$ | N̈ | $\stackrel{\sim}{\sim}$ | $\underset{\sim}{\underset{N}{N}}$ | $\stackrel{\sim}{\sim}$ | $\begin{gathered} 0 \\ \stackrel{\sim}{\sim} \end{gathered}$ | ̇ㅜN | $\stackrel{\infty}{\sim}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 3 | 2 | 9 | 6 | 5 | － | 5 | 2 | － | － | ＋ | ＋ | － | ＋ | － | － | － | － | － | － | － | － | － | － | － | － | －－ | － |
| Denmark | 1 | 3 | 4 | 5 | 10 | 4 | 1 | 2 | 1 | ＋ | ＋ | － | ＋ | ＋ | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Faroe Is． | － | 2 | － | － | － | 15 | 4 | 2 | 2 | － | 1 | － | － | 2 | 2 | 3 | 2 | 2 | 6 | 12 | 1 | 5 | － | 1 | ＋ | 1 | ＋ | － |
| France | 1910 | 2308 | 2496 | 2382 | 2648 | 2899 | 2059 | 1635 | 1814 | 1180 | 1135 | 782 | 1403 | 1366 | 1689＾ | 1537 | 2090 | 2073 | 1852 | 1374 | 1676 | 1622 | 1777 | 1246 | 1326 | 1734 | 1882 | 1287 |
| Germany | 1 | 2 | 163 | 140 | 160 | 113 | 249 | 269 | 194 | 158 | 78 | 38 | 91 | 105 | 116 | 73 | 222 | 146 | 211 | 166 | 149 | 142 | 136 | 151 | 201 | 258 | 246 | 394 |
| Ireland | 522 | 820 | 524 | 438 | 853 | 807 | 764 | 879 | 692 | 596 | 609 | 355 | 348 | 232 | 391 | 445 | 540 | 371 | 419 | 617 | 596 | 581 | 572 | 572 | 602 | 741 | 793 | 878 |
| Nether－ <br> lands | － | － | － | － | － | － | 27 | 1 | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Norway | 24 | 24 | 25 | 30 | 18 | 15 | 5 | 9 | 6 | 14 | 7 | 4 | 6 | 5 | 5 | 7 | 8 | 7 | 9 | 14 | 7 | 6 | 10 | 4 | 8 | 12 | 5 | 4 |
| Portugal | － | － | － | 132 | 128 | － | 91 | 413 | 429 | 20 | 18 | 8 | 4 | 19 | 63 | － | － | － | － | － | － | － | － | － | － | － |  | － |
| Russia | － | － | － | － | － | － | － | － | － | － | 1 | － | － | 2 | 4 | 1 | 1 | 35 | － | － | － | － | － | 1 | 2 | － | 2 | － |
| Spain | 340 | 274 | 186 | 215 | 333 | 229 | 234 | 338 | 344 | 231 | 397 | 229 | 255 | 153 | 117＾ | 112 | 15 | 259 | 242 | 229 | 167 | 105 | 123 | 81 | 149 | 234 | 290 | 261 |
| UK（E，W\＆NI） | 369 | 524 | 299 | 420 | 425 | 345 | 403 | 307 | 171 | 316 | 237 | 165 | 164 | 84 | 113 | 70 | 188 | 6 | 60 | － | － | 132 | 401 | － | － | － | － | － |
| UK（Scot） | 2814 | 2609 | 2528 | 2414 | 2732 | 2583 | 2478 | 1962 | 2032 | 1870 | 1486 | 1417 | 865 | 1156 | 1482 | 1451 | 1546 | 1720 | 2005 | － | － | 2073 | 1770 | － | － | － | － | － |
| UK（total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | 2234 | 2145 | 2205 | 2171 | 2310 | 2724 | 3080 | 3310 | 3387 |
| Total | 5984 | 6568 | 6234 | 6182 | 7312 | 7010 | 6320 | 5819 | 5685 | 4385 | 3969 | 2998 | 3136 | 3124 | 3982 | 3700 | 4613 | 4619 | 4804 | 4645 | 4742 | 4666 | 4790 | 4365 | 5013 | 6060 | 6528 | 6211 |
| Unallocated | 296 | 2638 | 3816 | 2634 | 4984 | 11148 | 7415 | 4821 | 3790 | 3131 | 1890 | 22 | 68 | 6 | 235 | 209 | 137 | 228 | 388 | 733 | 49 | 294 | 63 | 1515 | 5 | 94 | 68 | －145 |
| As used by WG | 6280 | 9206 | 10050 | 8816 | 12296 | 18158 | 13735 | 10640 | 9475 | 7516 | 5859 | 2976 | 3068 | 3130 | 3747 | 3491 | 4476 | 4847 | 5192 | 3912 | 4693 | 4372 | 4727 | 5880 | 5008 | 5966 | 6460 | 6356 |

## Table 4.4. Nominal landings (t) of Anglerfish in the North Sea, as officially reported to ICES.

## Northern North Sea (4.a). *Preliminary.

|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 2 | 9 | 3 | 3 | 2 | 8 | 4 | 1 | 5 | 12 | - | 8 | 1 | - | - | - | - | - | - | - | - | - | + | - | - | - | 1 | - |
| Den- <br> mark | 1245 | 1265 | 946 | 1157 | 732 | 1239 | 1155 | 1024 | 1128 | 1087 | 1289 | 1308 | 1523 | 1538 | 1379 | 1311 | 961 | 1071 | 1134 | 1143 | 841 | 821 | 854 | 801 | 962 | 1506 | 2002 | 1790 |
| Faroes | 1 | - | 10 | 18 | 20 | - | 15 | 10 | 6 | - | 2 | - | 3 | 11 | 22 | 2 | - | - | 4 | - | - | - | - | - | - | - | - | - |
| France | 124 | 151 | 69 | 28 | 18 | 7 | 7 | 3 | 18 | 8 | 9 | 8 | 8 | 8 | 4 | 7 | 13 | 13 | 20 | 23 | 20 | 14 | 15 | 27 | 26 | 35 | 91 | 141 |
| Germany | 71 | 68 | 100 | 84 | 613 | 292 | 601 | 873 | 454 | 182 | 95 | 95 | 65 | 20 | 84 | 173 | 186 | 344 | 216 | 124 | 46 | 265 | 274 | 321 | 286 | 208 | 523 | 510 |
| Netherlands | 23 | 44 | 78 | 38 | 13 | 25 | 12 | - | 15 | 12 | 3 | 8 | 9 | 38 | 13 | 14 | 14 | 12 | 5 | 8 | 5 | 5 | - | 16 | - | 21 | 28 | 68 |
| Norway | 587 | 635 | 1224 | 1318 | 657 | 821 | 672 | 954 | 1219 | 1182 | 1212 | 928 | 769 | 999 | 880 | 1006 | 831 | 860 | 859 | 791 | 494 | 485 | 545 | 524 | 406 | 610 | 840 | 1230 |
| Spain | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | + | - | - | - | - | - |  |
| Sweden | 14 | 7 | 7 | 7 | 2 | 1 | 2 | 8 | 8 | 78 | 44 | 56 | 8 | 6 | 5 | 5 | 20 | 67 | - | - | - | - | - | - | 6 | 4 | 8 | 12 |
| UK(E, W\&NI) | 129 | 143 | 160 | 169 | 176 | 439 | 2174 | 668 | 781 | 218 | 183 | 98 | 104 | 83 | 34 | 99 | 303 | 13 | 320 | 371 | - | 248 | 550 | - | - | - | - | - |
| UK (Scotland) | 7039 | 7887 | 9712 | 11683 | 15658 | 22344 | 18783 | 13318 | 9710 | 9559 | 10024 | 8539 | 6033 | 6284 | 6003 | 7722 | 8304 | 8658 | 7509 | 5730 | - | 4622 | 4154 | - | - | - | - | - |
| UK (to- <br> tal) | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 6353 | 4870 | 4704 | 5943 | 8005 | 9296 | 10127 | 9638 |
| Total | 9235 | 10209 | 12309 | 14505 | 17891 | 25176 | 23425 | 16859 | 13344 | 12338 | 12861 | 11048 | 8523 | 8987 | 8424 | 10340 | 10632 | 11038 | 10067 | 8190 | 7760 | 6459 | 6393 | 7633 | 9690 | 11680 | 13620 | 13390 |

Table 4.4. Continued. Nominal landings ( $\mathbf{t}$ ) of Anglerfish in the North Sea, as officially reported to ICES.
Central North Sea (4.b). * Preliminary

|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 357 | 538 | 558 | 713 | 579 | 287 | 336 | 371 | 270 | 449 | 579 | 435 | 180 | 260 | 207 | 138 | 179 | 181 | 134 | 124 | 111 | 131 | 135 | 213 | 196 | 251 | 168 | 155 |
| Denmark | 345 | 421 | 346 | 350 | 295 | 225 | 334 | 432 | 368 | 260 | 251 | 255 | 191 | 274 | 237 | 276 | 173 | 237 | 248 | 194 | 286 | 301 | 192 | 334 | 369 | 584 | 565 | 411 |
| Faroes | - | - | 2 | - | - | - | - | - | - | - | - | 10 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| France | - | 1 | - | 2 | - | - | - | - | - | - | - | - | - | + | - | + | + | - | 3 | 6 | 2 | +- | +- | 1 | + | + | - | + |
| Germany | 4 | 2 | 13 | 15 | 10 | 9 | 18 | 19 | 9 | 14 | 9 | 17 | 11 | 11 | 9 | 14 | 12 | 22 | 17 | 21 | 17 | 10 | 10 | 17 | 23 | 18 | 14 | 26 |
| Ireland | - | - | - | - | - | - | - | - | - | - | - | - | 1 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Netherlands | 285 | 356 | 467 | 510 | 335 | 159 | 237 | 223 | 141 | 141 | 123 | 62 | 42 | 25 | 31 | 33 | 61 | 58 | 36 | 46 | 53 | 61 | 41 | 72 | 88 | 120 | 166 | 111 |
| Norway | 17 | 4 | 3 | 11 | 15 | 29 | 6 | 13 | 17 | 9 | 15 | 10 | 12 | 22 | 16 | 12 | 24 | 15 | 21 | 10 | 11 | 11 | 26 | 8 | 9 | 16 | 41 | 36 |
| Spain | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | + | - | - | - | - | - | - |
| Sweden | - | - | - | 3 | 2 | 1 | 3 | 3 | 4 | 3 | 2 | 9 | 2 | 1 | 4 | 4 | 6 | 9 | - | - | - | - | - | - | 3 | 7 | 10 | 12 |
| UK(E, W\&NI) | 669 | 998 | 1285 | 1277 | 919 | 662 | 664 | 603 | 364 | 423 | 475 | 236 | 167 | 120 | 96 | 108 | - | 105 | 85 | 88 | - | 85 | 70 | - | - | - | - | - |
| UK (Scotland) | 845 | 733 | 469 | 564 | 472 | 475 | 574 | 424 | 344 | 318 | 378 | 210 | 241 | 138 | 88 | 98 | - | 142 | 108 | 125 | - | 115 | 72 | - | - | - | - | - |
| UK (total) | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 293 | - | - | - | 284 | 200 | 142 | 175 | 297 | 201 | 143 | 72 |
| Total | 2522 | 3053 | 3143 | 3445 | 2627 | 1847 | 2172 | 2088 | 1517 | 1617 | 1832 | 1244 | 847 | 851 | 688 | 683 | 749 | 769 | 651 | 615 | 764 | 714 | 546 | 820 | 985 | 1196 | 1107 | 823 |

Table 4．4．Continued．Nominal landings（t）of Anglerfish in the North Sea as officially reported to ICES．
Southern North Sea（4．c）．＊Preliminary．

|  | ন্ন | N్ | $\underset{ন}{\text { ® }}$ | す | 윽 | இ\% | 人 | がㅇㄱ | 익 | O- | O-B | N | Ò N | ষ্ণ | 응 | Ò O | － | O웅 | OiO | 0 | $\underset{\sim}{-}$ | $\underset{\sim}{\sim}$ | $\underset{\sim}{n}$ | $\stackrel{\underset{N}{A}}{ }$ | $\stackrel{\sim}{\sim}$ | $\begin{aligned} & 0 \\ & \stackrel{1}{N} \end{aligned}$ | Ǹ | $\begin{aligned} & \stackrel{*}{\infty} \\ & \stackrel{\sim}{\sim} \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 13 | 12 | 34 | 37 | 26 | 28 | 17 | 17 | 11 | 15 | 15 | 16 | 9 | 5 | 4 | 3 | 3 | 4 | 6 | 7 | 6 | 2 | 2 | 4 | 5 | 2 | 1 | 1 |
| Denmark | 2 | ＋ | － | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ |  | ＋ | ＋ | － | － | ＋ | ＋ | ＋ | － | － | － | ＋ | － | ＋ |
| France | － | － | － | － | － | － | － | 10 | － | ＋ | － | ＋ | － | － | － | ＋ | ＋ | － | 1 | 1 | 1 | ＋ | ＋ | 1 | ＋ | 1 | ＋ | ＋ |
| Germany | － | － | ＋ | ＋ | ＋ | － | － | － | － | － | － | － | － | － | － | ＋ | － | ＋ | ＋ | － | ＋ | － | ＋ | ＋ | $+$ | ＋ | ＋ | ＋ |
| Netherlands | 5 | 10 | 14 | 20 | 15 | 17 | 11 | 15 | 10 | 15 | 6 | 5 | 1 | － | 1 | － | 1 | 1 | － | 2 | 1 | 1 | 1 | 19 | 10 | 8 | 5 | 8 |
| Norway | － | － | － | － | ＋ | － | － | － | ＋ | － |  | － | ＋ | － | － | ＋ | － | － | 1 | － | － | － | － | 1 | ＋ | － | － | 1 |
| UK（E\＆W\＆NI） | 6 | 17 | 18 | 136 | 361 | 256 | 131 | 36 | 3 | 1 | － | － | 10 | 3 | － | ＋ | － | ＋ | 1 | 1 | － | － | 1 | － | － | － | － | － |
| UK（Scotland） | ＋ | ＋ | ＋ | 17 | ＋ | 3 | 1 | ＋ | ＋ | ＋ | － | － | － | 7 | － | ＋ | － | － | － | － | － | － | － | － | － | － | － | － |
| UK（Total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | －＋ | －＋ | ＋ | 1 | 1 | 1 | ＋ | 1 | 2 | 1 | 1 | ＋ | 1 |
| Total | 26 | 39 | 66 | 210 | 402 | 304 | 160 | 78 | 24 | 31 | 21 | 21 | 20 | 15 | 5 | 3 | 4 | 5 | 8 | 11 | 8 | 4 | 5 | 27 | 16 | 11 | 7 | 11 |

Table 4．4．Continued．Nominal landings（t）of Anglerfish in the North Sea as officially reported to ICES．
Subarea 4 （North Sea）．
＊Preliminary．$/$＾indicates landings assigned to Subarea 4 but not to a division．

|  | İ | N | $\stackrel{\text { ®̈ }}{ }$ | オ্ন | $\stackrel{\text { ® }}{\text { ® }}$ | 。ٌ | $\stackrel{\text { ® }}{-1}$ | $\stackrel{\infty}{\circ}$ | 各 | 잉 | Ö | Ò | ò No | ষ্N | Oì | ర్సి | 人̀ N | : ì | Oి미 | $$ | $\underset{\sim}{7}$ | $\underset{\sim}{\sim}$ | $\stackrel{\sim}{\underset{\sim}{N}}$ | $\underset{\sim}{J}$ | $\stackrel{\Perp}{\sim}$ | ロٌ | $\stackrel{\underset{\sim}{i}}{ }$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 372 | 559 | 595 | 753 | 607 | 323 | 357 | 389 | 286 | 476 | 594 | 459 | 190 | 265 | 211 | 141 | 181 | 185 | 140 | 131 | 116 | 133 | 137 | 217 | 200 | 253 | 169 | 156 |
| Denmark | 1599＾ | 1686 | 1293＾ | 1509＾ | 1027 | 1464 | 1489 | 1456 | 1496 | 1347 | 1540 | 1563 | 1714 | 1812 | 1616 | 1587 | 1134 | 1308 | 1382 | 1337 | 1127 | 1122 | 1046 | 1135 | 1331 | 2090 | 2567 | 2201 |
| Faroes | 1 | － | 12 | 18 | 20 | － | 15 | 10 | 6 | － | 2 | 10 | 3 | 11 | 22 | 2 | － | － | 4 | － | － | － | － | － | － | － | － | － |
| France | 124 | 152 | 69 | 30 | 18 | 7 | 7 | 13 | 18 | 8 | 9 | 8 | 8 | 8 | 4 | 7 | 14 | 13 | 23 | 30 | 24 | 15 | 15 | 30 | 26 | 36 | 91 | 142 |
| Germany | 75 | 70 | 113 | 99 | 623 | 301 | 619 | 892 | 463 | 196 | 104 | 112 | 76 | 31 | 93 | 187 | 198 | 367 | 233 | 145 | 63 | 275 | 284 | 339 | 309 | 226 | 537 | 536 |
| Ireland | － | － | － | － | － | － | － | － | － | － | － | － | 1 | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Netherlands | 313 | 410 | 559 | 568 | 363 | 201 | 260 | 238 | 166 | 168 | 132 | 75 | 52 | 63 | 45 | 47 | 76 | 71 | 41 | 56 | 59 | 67 | 42 | 108 | 98 | 148 | 199 | 187 |
| Norway | 604 | 639 | 1227 | 1329 | 672 | 850 | 678 | 967 | 1236 | 1191 | 1227 | 938 | 781 | 1021 | 896 | 1018 | 855 | 875 | 881 | 802 | 505 | 496 | 572 | 533 | 415 | 626 | 881 | 1267 |
| Sweden | 14 | 7 | 7 | 10 | 4 | 2 | 5 | 11 | 12 | 81 | 46 | 65 | 10 | 7 | 9 | 10 | 26 | 76 | － | － | － | － | － | － | 10 | 11 | 18 | 25 |
| UK（E\＆W\＆NI） | 804 | 1158 | 1463 | 1582 | 1456 | 1357 | 2969 | 1307 | 1148 | 642 | 658 | 334 | 281 | 206 | 130 | 207 | 425 | 118 | 406 | 460 | － | 333 | 621 | － | － | － | － | － |
| UK（Scot－ land） | 7884 | 8620 | 10181 | 12264 | 16130 | 22822 | 19358 | 13743 | 10054 | 9877 | 10402 | 8749 | 6274 | 6429 | 6091 | 7820 | 8476 | 8800 | 7617 | 5855 | － | 4736 | 4226 | － | － | － | － | － |
| UK（Total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | 6638 | 5069 | 4847 | 6120 | 8303 | 9498 | 10270 | 9711 |
| Total | 11790 | 13301 | 15519 | 18162 | 20920 | 27327 | 25757 | 19026 | 14885 | 13986 | 14714 | 12313 | 9390 | 9853 | 9117 | 11026 | 11384 | 11813 | 10726 | 8815 | 8532 | 7177 | 6944 | 8481 | 10691 | 12887 | 14733 | 14225 |
| Unallocated | －1224 | －1573 | －2441 | －2732 | －5126 | 11087 | －7540 | －4999 | －3166 | －2422 | －2037 | 600 | 676 | 1330 | －579 | 1462 | 1561 | 1081 | 945 | 915 | 612 | 765 | 638 | 316 | 448 | 33 | 225 | －55 |
| WG estimate | 10566 | 11728 | 13078 | 15430 | 15794 | 16240 | 18217 | 14027 | 11719 | 11564 | 12677 | 11713 | 8714 | 8523 | 9696 | 9564 | 9823 | 10732 | 9781 | 7900 | 7920 | 6412 | 6306 | 8165 | 10243 | 12854 | 14508 | 14280 |

Table 4．5．Nominal landings（t）of Anglerfish in Division 3．a，as officially reported to ICES．

## ＊Preliminary．

|  | ন্ন | ন্ন্ন | $\stackrel{刃}{\sigma}$ | 茳 | 俞 | ঃ윽 | बু | 命 | 윽 | চ্ণী | O্N | O్N | 음 | ষ্ণ | ద్N | Ò | 这 | : | Oì | 음 | $\underset{\sim}{7}$ | N̈̀ | $\stackrel{\sim}{\sim}$ | $\underset{\sim}{J}$ | $\stackrel{n}{\sim}$ | $\stackrel{0}{\sim}$ | ̇ㅡㄹ | $\stackrel{*}{*}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 15 | 48 | 34 | 21 | 35 | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Denmark | 493 | 658 | 565 | 459 | 312 | 367 | 550 | 415 | 362 | 377 | 375 | 369 | 215 | 311 | 274 | 227 | 255 | 287 | 344 | 270 | 251 | 307 | 298 | 309 | 336 | 389 | 526 | 597 |
| France | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | ＋ | ＋ | － | － | ＋ | 1 | － |
| Germany | － | － | 1 | ＋ | － | 1 | 1 | 1 | 2 | 1 | － | 1 | － | 1 | 1 | 2 | 1 | 1 | 1 | 1 | 2 | 1 | 1 | － | 1 | 2 | 1 | 2 |
| Netherlands | － | － | － | － | － | － | － | － | － | － | － | － | 3 | 4 | 4 | 3 | 1 | 3 | － | 5 | － | － | － | 4 | 9 | 17 | 16 | 16 |
| Norway | 64 | 170 | 154 | 263 | 440 | 309 | 186 | 177 | 260 | 197 | 200 | 242 | 189 | 130 | 100 | 139 | 132 | 144 | 134 | 158 | 153 | 115 | 108 | 127 | 90 | 124 | 118 | 204 |
| Sweden | 23 | 62 | 89 | 68 | 36 | 25 | 39 | 33 | 36 | 27 | 46 | 55 | 71 | 73 | 79 | 54 | 44 | 51 | － | － | － | － | － | － | 42 | 53 | 81 | 95 |
| UK（Total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | ＋ | － | － |
| Total | 595 | 938 | 843 | 811 | 823 | 702 | 776 | 626 | 660 | 602 | 621 | 667 | 478 | 519 | 458 | 426 | 433 | 486 | 478 | 433 | 405 | 423 | 407 | 440 | 478 | 586 | 742 | 914 |
| Unallocated | － | － | － | － | － | － | － | － | － | － | － | 288 | 252 | 197 | 174 | 189 | 168 | 187 | 79 | 109 | 116 | 63 | 65 | 78 | 66 | －5 | －9 | －22 |
| As used by WG | － | － | － | － | － | － | － | － | － | － | － | 379 | 226 | 322 | 284 | 237 | 265 | 299 | 399 | 324 | 289 | 360 | 342 | 362 | 412 | 591 | 751 | 936 |

Table 4.6. Breakdown of WG estimates of commercial catches for 2017 and 2018 by main gear group and area.

| 2017 | 3.a |  | 4 |  | 6.a |  | 6.b |  | Total |  | \% of Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fleet | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards |
| Demersal trawl | 86 | 0.56 | 11402 | 128 | 4271 | 179 | 779 | 64 | 16538 | 372 | 76 | 49 |
| Nephrops trawl | 506 | 11 | 279 | 118 | 76 | 99 | 0 | 0 | 861 | 228 | 4 | 30 |
| Gillnets | 100 | 1 | 2179 | 45 | 107 | 7 | 639 | 52 | 3025 | 105 | 14 | 14 |
| Other/Not specified | 59 | 0.48 | 648 | 12 | 582 | 37 | 5 | 0.43 | 1294 | 50 | 6 | 7 |
| Total | 751 | 13 | 14508 | 303 | 5036 | 322 | 1423 | 116 | 21718 | 754 | 100 | 100 |
| 2018 |  |  |  |  |  |  |  |  |  |  |  |  |
| Fleet | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards |
| Demersal trawl | 78 | 2 | 10569 | 41 | 3501 | 69 | 982 | 35 | 15130 | 147 | 70 | 45 |
| Nephrops trawl | 566 | 28 | 142 | 40 | 84 | 19 | 0 | 0 | 792 | 87 | 4 | 27 |
| Gillnets | 198 | 9 | 2405 | 19 | 149 | 4 | 892 | 30 | 3644 | 62 | 17 | 19 |
| Other/Not specified | 94 | 4 | 1164 | 13 | 677 | 11 | 70 | 2 | 2005 | 30 | 9 | 9 |
| Total | 936 | 43 | 14280 | 113 | 4411 | 103 | 1944 | 67 | 21571 | 326 | 100 | 100 |

Table 4.7. Total biomass estimates with confidence intervals and relative standard errors from the 2005-2019 SIAMISS-Q2 surveys.

| Year | Biomass (t) | Confidence Interval |  | RSE | Percentage Biomass in subarea 4 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 38.617 | 23.479 | 53.755 | 20.0 | 48.27\% |
| 2006 | 40.985 | 34.478 | 47.492 | 8.1 | 53.49\% |
| 2007 | 50.392 | 43.676 | 57.108 | 6.8 | 56.62\% |
| 2008 | 53.546 | 42.421 | 64.671 | 10.6 | 55.51\% |
| 2009 | 38.060 | 32.987 | 43.133 | 6.8 | 44.82\% |
| 2010 | 42.279 | 30.429 | 54.129 | 14.3 | 51.90\% |
| 2011 | 33.254 | 24.846 | 41.662 | 12.9 | 44.96\% |
| 2012 | 36.325 | 29.704 | 42.946 | 9.3 | 41.59\% |
| 2013 | 38.395 | 31.020 | 45.770 | 9.8 | 37.04\% |
| 2014 | 52.884 | 42.769 | 62.999 | 5.2 | 40.25\% |
| 2015 | 67.915 | 58.782 | 77.047 | 6.9 | 43.66\% |
| 2016 | 77.946 | 66.831 | 89.060 | 7.275 | 56.39\% |
| 2017 | 87.896 | 74.222 | 101.569 | 7.937 | 53.47\% |
| 2018 | 77.661 | 66.258 | 89.064 | 7.491 | 37.80\% |
| 2019 | 58.575 | 46.189 | 70.962 | 10.789 | 40.49\% |

Table 4.8. Abundance and biomass estimates from the 2005-2019 SIAMISS-Q2 surveys by ICES subareas and divisions.

| ICES Subarea/Division | Abundance (millions) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| Subarea 4 (partial) | 11.168 | 12.844 | 15.304 | 12.613 | 8.279 | 7.366 | 5.15 | 5.432 | 8.470 | 17.553 | 18.266 | 21.666 | 23.691 | 11.819 | 14.606 |
| Division 4.a | 10.866 | 10.459 | 7.956 | 7.718 | 5.144 | 5.161 | 6.057 | 4.961 | 8.461 | 16.096 | 28.604 | 14.383 | 16.322 | 13.528 | 21.032 |
| Division 4.b | 1.8 | 3.174 | 4 | 3.952 | 3.688 | 3.131 | 3.669 | 5.135 | 4.885 | 6.488 | 5.496 | 4.538 | 4.36 | 6.240 | 3.592 |
| Subarea 6 | 12.666 | 13.633 | 11.956 | 11.67 | 8.832 | 8.292 | 9.725 | 10.096 | 13.346 | 22.584 | 34.100 | 18.922 | 20.682 | 19.768 | 24.624 |
| Northern Shelf (partial) | 23.833 | 26.477 | 27.261 | 24.283 | 17.111 | 15.658 | 14.875 | 15.528 | 21.816 | 40.136 | 52.366 | 40.569 | 44.373 | 31.586 | 39.586 |
|  | Biomass (kilo tonnes) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| Subarea 4 (partial) | 18.642 | 21.921 | 28.534 | 29.721 | 17.058 | 21.944 | 14.949 | 15.106 | 14.369 | 21.284 | 29.653 | 43.956 | 46.994 | 29.353 | 23.353 |
| Division 6.a | 14.096 | 12.175 | 11.072 | 14.383 | 8.15 | 11.59 | 9.33 | 9.213 | 10.801 | 16.633 | 24.047 | 18.273 | 29.296 | 22.350 | 18.864 |
| Division 6.b | 5.879 | 6.889 | 10.786 | 9.442 | 12.852 | 8.745 | 8.974 | 12.005 | 13.626 | 14.967 | 14.215 | 15.717 | 11.604 | 25.958 | 15.992 |
| Subarea 6 | 19.975 | 19.064 | 21.858 | 23.825 | 21.002 | 20.334 | 18.305 | 21.218 | 24.427 | 31.600 | 38.262 | 33.990 | 40.901 | 48.308 | 34.856 |
| Northern Shelf (partial) | 38.617 | 40.985 | 50.392 | 53.546 | 38.06 | 42.279 | 33.254 | 36.325 | 38.796 | 52.884 | 67.915 | 77.946 | 87.896 | 77.661 | 58.575 |

Table 4.9. Percentage change in mean stock biomass from 2015-2017 to 2018-2019 in ICES subareas 4 and 6 combined.

| Average Biomass 2015-2017 | Average Biomass 2018-2019 Percentage Change in Biomass |  |
| :---: | :---: | :---: |
| 77.919 | 68.118 | $12.578 \%$ |



Figure 4.1. Northern Shelf anglerfish. Officially reported landings by ICES area (1973-2018).


Figure 4.2. Trends in nominal international fishing effort ( $\mathbf{k W}^{*}$ days at sea) in North Sea and II (EU) (left) and West of Scotland (right) collated by STECF for the Evaluation of Fishing Effort Regimes in European Waters (STECF, 2017).
anf.27.3a46 27.4 LandPercent


Figure 4.3a. Percentage of total landings weight by fleet and country in 2018; Subarea 4.


Figure 4.3b. Percentage of total landings weight by fleet and country in 2018; Division 6.a.
anf.27.3a46 27.6.b LandPercent


Figure 4.3c. Percentage of landings weight by fleet and country in 2018; Division 6.b.


Figure 4.4. WGCSE Landed numbers ('00 thousands) at-length (cm) 2002-2018.


Figure 4.5. SIAMISS-Q2 estimates of total biomass, with confidence intervals, for subareas 4 and 6 combined, 2005-2019. B now is the average biomass for 2018-2019, $B_{\text {ref }}$ is the average biomass for 2015-2017; both marked on the graph in their respective years. Ratio Est is the ratio of $B_{\text {now }}$ to $B_{\text {ref, }}$ expressed as a percentage, with confidence intervals (Ratio CIlo, Ratio CIup).


Figure 4.6. SIAMISS-Q2 estimates of total abundance (left) and biomass (right) of anglerfish for the Northern Shelf (black filled squares) 2005-2019. Estimates are also provided for ICES Subarea 4 (blue filled circles), Division 6.a (red triangles) and Division 6.b (turquoise diamonds).


Figure 4.7. SIAMISS-Q2 estimates of total numbers (millions) at-length (cm) for subareas 4.a-c and 6.a-b, 2019.


Figure 4.8. SIAMISS-Q2 estimates of total numbers (millions) at-length (cm) (left) and estimates of total biomass (kt) at-length (cm) (right) for subareas 4.a-c and 6.a-b combined, 2007-2019.


Figure 4.10. Percentage of SIAMISS-Q2 total biomass, with confidence intervals, estimated to be in subareas 4.a-c compared with subareas $4 . a-c$ and $6 . a-b$ combined. The full grey line represents the average of these percentages over the time-series (2005-2019) 4 (47\%). The dotted grey lines represent the percentage of TAC allocated for subareas $4 . a-c$ compared to the total of the TAC for subareas $4 . a-c$ and $6 . a-b,(67 \%$ in 2005-2010, 64\% in 2011-2017).


Figure 4.11. Northern Shelf anglerfish harvest rate 2008-2018 (mean standardised WG catch total numbers of biomass)/ SIAMISS-Q2 total numbers or biomass).


Figure 4.12. Mean length for NS-IBTS-Q1 (black bold), NS-IBTS-Q3 (black), commercial catch WKAnglerfish InterCatch estimation (red), and SIAMISS-Q2 raw catches (blue crosses) and after survey estimation procedure of corrections for footrope escapes, herding etc. (blue open circles) (see Reid et al., 2007).


Figure 4.13. Survey indices of mean weight $(\mathrm{g})$ per hour from SWC-IBTS-Q1 (blue) in 6.a, SWC-IBTS-Q4 (red) in 6.a and ROCKALL (green) in 6.b.

Anglerfish
North Sea Q1 and Q3 - all countries


Figure 4.14. Survey indices of mean weight (g) per hour from NS-IBTS-Q1 (brown) and NS-IBTS-Q3 (orange).

## 3 Cod in Division 6.a

### 3.1 Introduction

The last benchmark for this stock was carried out in February 2012 (ICES, 2012) with subsequent inter-benchmarks in February 2015 (ICES, 2015) and February 2019 (ICES, 2019). The assessment and forecast carried out at the WG follows the procedure outlined in the stock annex developed at the benchmark and updated at the inter-benchmarks, with the exception of some minor deviations:
i. additional down-weighting of individual datapoints to improve TSA assessment model diagnostics (the stock annex acknowledges the need to allow for changes to the variance structures used in the TSA models if they improve model diagnostics);
ii. The forecast assumptions differ from those used at previous assessment WGs and those documented in the Stock Annex which have not been discussed or modified since 2008.

The deviations are explained in the relevant report section.

### 3.2 General

### 3.2.1 Advice

Biennial advice was provided for this stock in 2017. This, and advice from previous years is given below.

ICES Advice applicable for 2018 and 2019
ICES advises that when the MSY approach is applied, there should be zero catches in each of the years 2018 and 2019.

## ICES Advice applicable for 2016 and 2017

ICES advises that when the MSY approach is applied, there should be no directed fisheries and all catches should be minimized in 2016 and 2017.

## ICES Advice applicable for 2015

No new data are available that change the perception of the stock from the advice given in 2013. Therefore, the same catch advice is still applicable for 2015: ICES advises on the basis of the MSY and precautionary approach that there should be no directed fisheries and that bycatch and discards should be minimized.

### 3.2.2 Stock definition and the management unit

The assessment unit is Division 6.a although there are believed to be at least two subpopulations of cod in Division 6.a which remain geographically separated throughout the year. Further details can be found in the stock annex. The management unit is ICES Divisions 6.a plus EU and international waters of Division $5 . b$ to the east of $12^{\circ} 00^{\prime} \mathrm{W}$. Prior to 2009, the TAC was set for ICES Subareas 6, 12 and 14 plus Subdivision 5.b.1.

## Recent management

The minimum conservation reference size of cod for human consumption in this area is 35 cm .
From 2012 to 2018 the TAC for cod in Division 6.a was set to zero with allowance for a bycatch of cod to be landed provided that it does not comprise more than $1.5 \%$ of the live weight of the total catch retained on board per fishing trip. From 2015, this provision has not been allowed for catches subject to the landing obligation.

With the full implementation of the landing obligation in 2019 for fisheries catching cod, a bycatch TAC of 1735 t has been set to allow mixed fisheries with a bycatch of cod to continue.

## TAC for 2012-2014

| Species: | Cod <br> Gadus morhua | Zone:Vla; Union and international waters of Vb east <br> of $12^{\circ} 00^{\prime} \mathrm{W}$ <br> $(\mathrm{COD} / 5 \mathrm{BE6A})$ |
| :--- | :--- | :--- | :--- |
| Belgium | 0 |  |
| Germany | 0 |  |
| France | 0 |  |
| Ireland | 0 |  |
| United Kingdom | 0 | Analytical TAC |
| Union | $\left.0{ }^{1}\right)$ |  |
| TAC |  |  |

$\left.{ }^{( }{ }^{1}\right)$ By-catch of cod in the area covered by this TAC may be landed provided that it does not comprise more than $1,5 \%$ of the live weight of the total catch retained on board per fishing trip.

## TAC for 2015-2018

| Species:Cod <br> Gadus morhua | Zone:Vla; Union and international waters of Vb east of <br> $12^{\circ} 00^{\prime} \mathrm{W}$ <br> $(\mathrm{COD} / 5 \mathrm{BE} 6 \mathrm{~A})$ |  |  |
| :--- | :--- | :--- | :--- |
| Belgium | 0 |  |  |
| Germany | 0 |  |  |
| France | 0 |  |  |
| Ireland | 0 |  |  |
| United Kingdom | 0 |  |  |
| Union | 0 | $\left.{ }^{1}\right)$ | Analytical TAC |
| TAC |  |  |  |

[^2]TAC 2019

| Species: Cod <br>  Gadus morhua |  | Zone: | 6a; Union and international waters of 5 b east of $12^{\circ} 00^{\prime} \mathrm{W}$ <br> (COD/5BE6A) |
| :---: | :---: | :---: | :---: |
| Belgium | $3{ }^{1}{ }^{1}$ |  |  |
| Germany | 26 (1) |  |  |
| France | $275{ }^{(1)}$ |  |  |
| Ireland | $385{ }^{(1)}$ |  |  |
| United Kingdom | $1046{ }^{1}{ }^{1}$ |  |  |
| Union | $1735{ }^{(1)}$ |  |  |
| TAC | $1735{ }^{(1)}$ |  | Analytical TAC <br> Article 8 of this Regulation applies |

$\left.{ }^{( }{ }^{1}\right)$ Exclusively for by-catches of cod in fisheries for other species. No directed fisheries for cod are permitted under this quota.

Technical measures applicable to the West of Scotland, including those associated with the cod recovery plan in force up to 2008 (Council Regulation No. 423/2004), the cod long-term management plan in force from 2009 (Council Regulation No. 1342/2008) were amended by Council Regulation No. 1243/2012. The management plan was further amended in 2016 by Council Regulation (EU) 2016/2094 to cover the transitional period in which preparations are ongoing towards multiannual plans for multispecies fisheries. In 2018 the cod management plan was discontinued. Cod in Division 6.a is not included in the multiannual plan for Western Waters (Council Regulation (EU) 2019/472).

### 3.2.3 The fishery in 2018

The table of official landings statistics is given in Table 5.1. Official landings have increased in recent years and in 2018 were 360 tonnes, more than double the 2014 value ( $\sim 160$ tonnes), which was the lowest of the time-series. In 2018, almost $60 \%$ of the official landings were reported by UK vessels, approximately $30 \%$ by France with smaller amounts declared by Norway, Ireland, Denmark and the Faroe Islands. The majority of reported cod landings in Division 6.a are now taken in the far north of the area (Figure 5.1 shows Scottish reported landings by statistical rectangle).
Due to restrictive TACs, seasonal/spatial closures of the fishery, and effort restrictions based on bycatch composition, the likelihood of misreporting and underreporting of cod in the past is considered to have been high. Underreporting is considered to have been reduced to low levels following the introduction of legislation in Ireland and the UK in 2006. However, area misreporting of cod landings from Division 6.a into Division $4 . a$ (i.e. caught in Division 6.a., but declared in Division 4.a) and to a lesser extent Division 5.b, by the Scottish fleet is now believed to occur. The UK legislation introduced in 2006 is also believed to be responsible for a significant increase in discards starting in 2006.

Area-misreported landings by the Scottish fleet are considered to represent a considerable proportion of the total landings. Estimates of misreporting based on surveillance and consideration of VMS data by Marine Scotland Compliance, have been made available to the WG. Figure 5.2 shows the time-series of misreporting estimates which are assumed to come from the large mesh demersal trawl fleet. Total estimated area misreported Division 6.a cod landings in 2018 were 741 t (largely reported into Division 4.a and to a lesser extent 5.b). This represents over $65 \%$ of the total landings in 2018.

### 3.3 Data

## Catch data

The landings uploaded into InterCatch are shown in Figure 5.3 by métier and country, and discard weights and proportions are shown in Figures 5.4 and 5.5 respectively. With the exception of the area misreported landings, the French OTB_DEF $\geq 120$ métier is the largest metier with unsampled (no age-compositions) landings ( $\sim 9 \%$ of the total landings in 2018)

There are no age composition samples from the misreported landings. The WG this year followed the same procedure as the last two years for handling the misreported landings within InterCatch (a deviation from the Stock Annex). Previously, landings numbers-at-age from the Scottish demersal fleet (OTB_DEF>=120) were raised to the total reported plus area-misreported landings prior to uploading to InterCatch. However, the 'misreporting fleet' could potentially have a different landings age composition (as they are assumed not to discard) and the 2017 WG considered that a more appropriate approach would be to upload the misreported landings into InterCatch as a separate unsampled fleet. This allows a weighted average landings age composition (Irish and Scottish) to be applied. (The Irish landings comprise a substantially greater proportion of younger fisher than the Scottish sampled landings, although given the relative landings weights of the two fleets, the allocated proportions are similar to the Scottish sampled fleet).

It can be seen that landings by Scottish trawl $\geq 120 \mathrm{~mm}$ dominate, and total discards are also highest from this fleet (Figures 5.3 and 5.4). However, the discard proportion is higher for the Scottish trawl 70-100 mm fleet (OTB_CRU_70-99) (Figure 5.5). The discard rates observed in the Irish and French fleets are considerably lower than in the Scottish $\geq 120 \mathrm{~mm}$ demersal fleet. The proportions of the catch discarded (by weight) for the sampled fleets are given below.

| Fleet | Scottish Demersal <br> Trawl^ | Scottish Nephrops <br> Trawl | Irish Demersal | French Demersal <br> trawl |
| :--- | :--- | :--- | :--- | :--- |
| Discard proportion | $73 \%$ | $96 \%$ | $24 \%$ | $38 \%$ |

${ }^{\wedge}$ The calculation of this discard proportion excludes the area misreported component of landings as this fleet are assumed not to discard i.e. this is the discard rate of the 'reported landings fleet'.

Discard proportions and landings and discard age distributions were assigned within InterCatch to unsampled fleets on the same basis where possible (and as described in the Stock Annex). Raised discards are shown in Figure 5.6. The final mix of numbers-at-age from sampled and unsampled landings and sampled and raised (unsampled) discards is given in Figure 5.7. An extremely small amount ( 35 kg ) of below minimum size (BMS) landings was also reported, but is not shown. The large unsampled proportion of the catch-at-ages three and above is due to the landings from the Scottish misreported fleet.

Sampling levels (number of trips) by country are given below. A limited number of Northern Irish samples are also available in some years. Sampling of the Scottish OTB_DEF landings is still relatively poor. The small sample sizes (which include a few very large fish with high raising factors) can result in a very high sum of products (SOP, landings-at-age x weight-at-age) for this fleet in some years (2015 and 2016).

|  | Scotland |  | Ireland |  |
| :--- | :---: | :---: | :---: | :---: |
| Year | Demersal trawl (OTB_DEF) | Nephropstrawl (OTB_CRU) | Total | Total |
| Landings | 11 | 1 | 12 | 24 |
| Observer | 11 | 28 | 39 | 12 |

The WG estimates of total landings and discards are given in Table 5.2 and shown in Figure 5.8. These values are for fish aged 1 to $7+$ which is the age range used in the assessment. Just over three tonnes of age zero fish were also estimated to have been discarded in 2018.

The total discard proportion by weight is shown in Figure 5.9. The estimate of total discards as a proportion of total catch by weight in $2018(40 \%)$ is the lowest since 2005. However, this reduction is mostly due to the increase in the estimate of area misreported landings for 2018 which are assumed to have zero discard rate. Furthermore, the discard estimates are highly uncertain; the CV of the discard weight estimate for the large mesh Scottish demersal trawl fleet for 2018 is $51 \%$. Given the $1.5 \%$ bycatch regulation, the landings have potentially been limited more by catch-rates of other species in the fishery. So, for example, an increase in the catch rate of anglerfish and/or haddock could have allowed for a greater proportion of cod catches to be landed by the Scottish demersal fleet in these years.

Discarding occurs across most of the age classes in the catch (Figure 5.10).

## Age compositions

Raised landings numbers-at-age and discard numbers-at-age are given in Tables 5.4 and 5.6 respectively, and total catch numbers-at-age in Table 5.8. In 2018 there has been a large reduction in the catches-at-ages 1 to 3 compared to 2017. (Figure 5.11).

## Weight-at-age

Annual mean weights-at-age in landings, discards and catch are given in Tables 5.5, 5.7 and 5.9. Figure 5.12 shows the mean weights-at-age in the landings and discards. The mean weight of age two and three fish in the landings has increased since the mid-2000s in line with the increase in high-grading which has occurred at these ages. Other age classes show fluctuations with a long-term downward trend particularly for ages 5 and above. Values at older age are noisy, particularly in recent years (most likely due to low sampling levels). Mean weight-at-age in the discards shows no real trend.

## Survey data

All available survey data are given in Table 5.3, with the data used in the assessment highlighted in bold. Survey descriptions are given in the stock annex. Following the inter-benchmark (IBPCod6.a), the assessment now makes use of three additional quarter four surveys (one of which is no longer current). Survey indices for the two new Scottish surveys (UK-SCOWCGFS- Q1 and UK-SCOWCGFS- Q4) are provided with an estimate of variance.

The cpue by survey haul for the IRGFS-WIBTS-Q4 survey are shown in Figure 5.13 and in Figure 5.14 for the two Scottish surveys (UK-SCOWCGFS- Q1 and UK-SCOWCGFS- Q4). All surveys show mostly zero returns over latitudes between 56 degrees N and 58.5 degrees N (although the IRGFS-WIBTS-Q4 survey only extends to 56.5 degrees N). This pattern has been consistent in
surveys since 2007. The Scottish surveys have highest catch rates to the north of 59 degrees N , in and around the 'windsock' closed area. South of 56 degrees N , the Q1 surveys catch cod in the Clyde region and the Q4 surveys catch some cod off the Northern Irish coast. From the IRGFS-WIBTS-Q4 survey there is also evidence of higher abundance in this area as well as along the shelf edge in the southern part of Division 6.a. Catch rates of age one cod are typically very low and the higher catches of older age classes that appear in the north of the region could potentially be due to overspill from the neighbouring North Sea stock which has increased in recent years. The UK-SCOWCGFS-Q1 in 2019 shows very low catch rates for ages $>1$ across the area, but relatively high catch rates (compared to recent years) of age 1 fish.

In 2017, the indices for age four, five and six cod in the quarter one survey show particularly high uncertainty due to a single very large haul (Figure 5.14) of large cod with most other stations having very low or zero values. In 2018, there were no large hauls and therefore the estimated variance is low. In 2019, the quarter one survey shows very low catch rates of ages $>1$ across the survey area, but relatively high catch rates (compared to recent years) of age 1 fish.

The quarter four survey estimates also have substantial uncertainty. This is particularly apparent in the 2018 survey with two hauls catching large numbers of individuals aged 4 to 6 and very low catches elsewhere, resulting in CVs of around $60 \%$ for these ages in this year.

A series of inshore and offshore Scottish industry-science surveys, known as the West Coast Demersal Fish (WCDF) project were conducted between December 2013 and November 2014. The initiative, funded by the Scottish Government and the European Fisheries Fund, was a joint venture between Marine Scotland Science and the Scottish Fishermen's Federation with the aim of improving the understanding of the current state of demersal stocks to the West of Scotland. The surveys show a broadly similar distribution to the UK-SCOWCGFS- Q1 and UK-SCOWCGFS-Q4 with bigger fish and increased abundance inside the Windsock compared to outside.

## Biological data

Natural mortality-at-age (M) is assumed to be weight-dependent after Lorenzen (1996) but time invariant. M is calculated by finding the time-series means for stock weights-at-age before applying the Lorenzen parameters and the values are shown below.

Natural mortality (M) at-age:

| Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | 7+ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 0.540 | 0.387 | 0.307 | 0.263 | 0.238 | 0.224 | 0.212 |

Figure 5.15 shows the resulting M-at-age values used in the assessment (the constant values) and the values calculated based on the annual mean weights-at-age for comparison. This suggests that in future it may be more appropriate to include time varying natural mortality given the changes in mean weight-at-age that have occurred. Proportion of fish mature-at-age are unchanged from the last meeting and is as detailed in the stock annex.
The contribution of seal predation to total cod mortality is likely to be significant and this may impair the ability of the stock to recover (Cook et al., 2015). Weight dependent natural mortali-ties-at-age were adopted at the benchmark meeting in 2012 to take account of increased mortality at younger ages. Given the sparsity of the data, it is considered unlikely that seal consumption estimates could be incorporated into a robust annual stock assessment.

### 3.4 Stock assessment

This assessment uses a TSA run as outlined in the stock annex. Exploratory analysis of the input catch and survey data are also carried out.

## Data screening

Catch curves from commercial catch-at-age data (landings plus discards) are shown in Figure 5.16. Although the data are noisy, there is some evidence of a flattening off of the catch curves in recent years compared to those of the cohorts spawned in the late 1990s. A plot of log catch curve gradients derived from commercial catch data (landings plus discards) over different age ranges is shown in Figure 5.17. Here too there is some evidence of a decreasing mortality in recent years, particularly over age ranges including age two. (Note that these exploratory catch data plots are based on reported landings and discards and will be influenced in part by underreporting of landings in the 1990s and early 2000s).

Figure 5.18 shows the mean standardised catch-at-age by proportion (number). It shows good tracking of the strong cohorts as recently as the 2005 year class which shows well even at age 7+. More recently the data become rather noisy and in 2015, the proportion of the catch-at-age six is high and similarly for age $7+$ in 2016. Neither of these observations are supported by above average values at younger ages of the same cohort. Potentially the age 6 value in 2015 and $7+$ in 2016 could be an overspill of fish from the North Sea as this coincides with the moderately strong 2009 year class in that area. In 2019, a small increase in the numbers-at-age four and older in the catch, coincidental with a large reduction in the numbers-at-age one to three results in the proportion at older ages in the catch being the highest of the time-series (Figure 5.18).

Figure 5.19 shows the log mean standardised indices from the ScoGFS-WIBTS-Q1 survey by year and by cohort. The early part of the time-series appears to track the cohorts relatively well with no obvious year effects. However in later years, the indices become more noisy and there is some evidence of year effects in the survey. The survey ended in 2010.

Figure 5.20 shows log catch curves for the ScoGFS-WIBTS-Q1 survey. It shows a strong "hook" at the younger ages, with abundance-at-age two often higher than at-age one. In later years survey abundance also shows increases from age 2 to age 3 in the same year class and the survey's ability to track recent cohorts seems poor relative to the 1990s and early 2000s. The survey scatterplots (Figure 5.21) show some consistency in the estimates of year-class strength across age classes (particularly the younger, adjacent ages), although less so at older ages. There is no trend in the log catch curve gradients derived from this survey that would be consistent with a change in mortality (Figure 5.22) for any of the age ranges considered.

Figure 5.23 shows the log mean standardised indices by cohort and year from the ScoGFS-WIBTS-Q4 survey. The survey shows reasonable tracking of cohorts at ages one to three and no particular evidence of year effects. This is also evident in the survey scatterplots which show reasonable correlation at younger ages (Figure 3.24). This survey catches very few fish at ages five and above.

Figure 5.25 shows the log mean standardised indices by cohort and year from the IRGFS-WIBTSQ4. The $\log$ mean standardised indices plot shows consistent signals at ages 1 and 2 with no obvious year effects. The scatterplots (Figure 5.27) also show reasonable consistency between ages one and two, but the tracking at older ages is less strong. The data cover too few age classes sufficiently well to give an indication of trend in mortality through catch curve gradients (Figure 5.26).

Figure 5.28 shows log mean standardised indices by cohort and year from the UK-SCOWCGFSQ1. There is little evidence of successful tracking of cohorts and some evidence of survey year effects (2015, 2017 and 2019, particularly for older ages). There appeared to be a general increase in the catch rates of older ages over time to 2017 (four and above), but no equivalent increase in the catch rates of younger ages (from the same cohort). These declined significantly in 2018 and 2019 although there has been an increase in the catch rate of age one in 2019.

The log catch curves from the UK-SCOWCGFS- Q1 are also very noisy (Figure 5.29) and typically do not show a decline as the cohort ages. The survey scatterplots show that even the catch rates of successive age classes (within the same cohort) are only weakly related (Figure 5.30).
Figure 5.31 shows log mean standardised indices by cohort and year from the UK-SCOWCGFSQ4. There is some evidence of cohort tracking, but this is not consistent over time or ages and this is also apparent in the survey scatterplots shown in Figure 5.33. Figure 5.32 shows the log catch curves from the UK-SCOWCGFS-Q4, which are noisy and difficult to interpret given the short time-series and missing year of survey data.
Overall, information on mortality trends from all survey-series (including the ScoGFS-WIBTSQ1) appears to be fairly poor due to the generally high variability and large CVs (ranging from $30 \%$ to $75 \%$ depending on age class) for the two current Scottish surveys.

Figure 5.34 shows a comparison (between surveys) of log mean standardised survey indices atage over time (mean standardised over the common year range of all three surveys). The two quarter four surveys show some consistency over time at-age two while the two Scottish surveys show some consistency of trends at-age three. At older ages (in the Scottish surveys), there appears to be a divergence in the trend in recent years.

The inter-benchmark in 2019 agreed that all five surveys should be included in the final assessment, the basis being that the additional surveys show reasonable internal consistency and in addition, some between survey consistency. It was considered that the Irish survey could provide an additional indicator of year-class strength and could be useful as it covers the period during which there is a break in the Scottish survey indices. The lack of spatial coverage of this survey (only the southern part of Division 6.a) was deemed less important given the index is only being used to provide information on the younger ages.

## Final assessment

Model settings and input parameter settings for the final run are given in Table 5.10. Input data are as agreed at the 2019 inter-benchmark and include the five surveys described above, and commercial landings- and discards-at-age (age compositions only from 1991-2005, and data excluded completely for 2006). Final parameter estimates from the TSA run are given in Table 5.11. Running the update assessment from the 2019 inter-benchmark identified a large positive residual in age one landings in 2019 in the model diagnostics (Figure 5.35), due to increased landings and lower discards at-age 1 (i.e. a lower discard rate than the recent past, which the model is unable to replicate). The only observations of age 1 landings were in the Irish sampled data (i.e. not in the Scottish) and therefore it was considered appropriate to allow additional uncertainty for this datapoint to improve model diagnostics. An alternative model run in which the 2019 discard data were down-weighted made only limited improvement to the diagnostics. (Note that the stock annex acknowledges the need to allow for changes to the variance structures (down-weighting of individual datapoints) used in the TSA models if they improve model diagnostics).

The inter-benchmark (ICES, 2019) highlighted the problems with the estimation of discards; the inability of the current discard model to reproduce the interannual variability in the discard proportions and also the influence of the proportion at-age one in recent years in the overall estimates of transitory variation in discarding. In future (requiring some bespoke changes to TSA), an assumption that all age one catch are discarded may provide a better fit to the full time-series of data (assuming that discarding practices do not change). Alternatively, an approach making use of an age-based discard ogive may be more appropriate than modelling proportions at-age as a random walk.

Figures 5.36 and 5.37 show the residuals by age class for landings and discards and the surveys respectively (based on the final smoothed model estimates). The landings and discards residuals are all reasonably small with no major outliers or particular patterns, with the exception of some indication of an underestimation of landings at-age one and overestimation of discards at-age one.

In terms of survey residuals, the two discontinued survey time-series continue to show some indication of increasing residuals over time for younger ages, most likely indicative of a mismatch in the commercial and survey data (although alternatively could be interpreted as an increase in survey catchability which is not accounted for in the model). The recent Scottish survey indices are very uncertain (and CVs are input to the assessment) while the catch data are still estimated to be comparatively more precise, resulting in a poor distribution of residuals for some ages (although the magnitude is low). Patterns are fairly consistent with fits in previous years and at the inter-benchmark.

The time-series of observed and fitted discard proportions-at-age is shown in Figure 5.38. The predictions follow the general trend in the data which are quite noisy. (See above for discussion on discard modelling).

Table 5.12 gives the TSA population numbers-at-age and Table 5.13 gives their associated standard errors. Estimated F at-age is given in Table 5.14 and standard errors of mortality on the log scale are given in Table 5.15. Full summary output is given in Table 5.16. A summary plot for this run is shown in Figure 5.39.

Retrospectives for the final assessment run are shown in Figure 5.40. Following discussion at WGCSE, the recruitment and SSB plots include the intermediate year in each of the peels (i.e. the latest assessment includes model output up to 2019 for R and SSB). There is little evidence of retrospective bias in either mean F or SSB, with respective Mohn's rho values of 0.11 and 0.03 (five year peel). Further back in time the mean $F$ shows some evidence of underestimation which appears to be associated with the break point in the Scottish survey indices (i.e. when the current survey series begin).

The assessment appears to overestimate recruitment in the intermediate year (the estimate is based on only a single survey datapoint). The recruitment Mohn's rho is 1.71 when the intermediate year of each assessment peel is included and 0.19 when the model estimates are limited to the last year of catch data.

## Stock status

Historical stock trends are shown in Figure 5.40 and the stock-recruitment relationship is shown in Figure 5.41. The estimated SSB shows a steady downward trend until 2006 and has fluctuated at a slightly higher level since then. The 2012 year class (recruitment in 2013) is estimated to be the highest since 2006 and results in a small increase in SSB in 2015 and 2016 (highest since 2003). Since then, recruitment has declined and in 2018 is estimated to be the lowest of the time-series.

Estimated SSB in the final year is well below Blim (= 14000 tonnes). Mean F is well above Fmsy, but has been below Flim since 2012. Although the latest assessment shows an increase in mean F since 2016, there has been a clear decrease in mean $F$ since 2005 . The decline in mean $F$ is proportionately similar ( $\sim 50 \%$ ) to the decline in STECF effort (large and small mesh demersal/crustacean trawl from both regulated and unregulated fleets), although the mean F does not start to decline until several years after the effort (Figure 5.42a). Partial mean F for landings and discards separately is shown in Figure 5.42b showing that discarding accounts for around $50 \%$ of the mean F.

### 3.5 Short-term stock projections

The inputs for the short-term forecast follow the specifications in the Stock Annex with the exception of the numbers-at-age one in 2019. The Stock Annex (which does not appear to have been updated since 2008 with respect to the forecast settings) states that the numbers-at-age at the start of the intermediate year ought to be the estimates from TSA. However, the retrospective plots discussed above suggest that the estimate of recruitment in the intermediate year from TSA is generally revised downwards in subsequent assessments. Therefore, in contrast to the stock annex (and previously presented forecasts) we use a short-term (ten year) geometric mean recruitment for the intermediate year onwards in the forecast.

Fishing mortality in the intermediate year (2019) was taken as a three year average over 2016 to 2018 of the exploitation pattern rescaled to the 2018 mean F as an estimate of F status quo (given that the mean F appears to be increasing). Mean weights-at-age were also averaged over the most recent three years.

In 2019 (the intermediate year), cod in Division 6.a is fully under the landing obligation and a bycatch TAC of 1735 t has been set to allow mixed fisheries with a bycatch of cod to continue. A change in discarding practices would be expected following this increase in TAC. Therefore, the forecast assumes that in 2019 and 2020, unwanted catch will be the catch below MCRS. Given that discards in recent years have been largely high grading (due to restrictive TACs), catch below MCRS is approximated by the use of historical discard proportions-at-age in the forecast (average over 1981-2000 i.e. before high-grading became an issue). This results in very low discard selectivity at all ages above age one.

The recent high-grading has resulted in landings mean weights-at-age significantly higher than catch weights-at-age. In the forecast, which assumes only below MCRS to be unwanted (i.e. no high-grading), wanted catch mean weights-at-age are assumed to be the three year average (2016-2018) of observed mean catch weights-at-age.

| Variable | Value | Notes |
| :--- | :---: | :--- |
| F ages 2-5 (2019) | 0.70 | Average exploitation pattern (2016-2018) scaled to F ages 2-5 in <br> 2018 |
| SSB (2020) | 2013 | Tonnes; short-term forecast. |
| Rage1 (2019 and 2020) | 2349 | Thousands; GM recruitment (2009-2018) |
| Catch (2019) | 1421 | Tonnes; short-term forecast. |
| Wanted catch (2019) | 1366 | Tonnes; assuming average discard proportions-at-age (1981- <br> 2000) |
| Unwanted catch (2019) | 55 | Tonnes; assuming average discard proportions-at-age (1981- <br> 2000) |

The short-term forecast inputs are shown in Table 5.17 and the outputs in Table 5.18 and 5.19. Note that the numbers-at-age in 2019 in Table 5.17 are the survivors from 2018 which differ slightly to the TSA numbers-at-age in 2019 (Table 5.12) which are smoothed estimates.

Under the forecast assumption of status quo F, wanted catch in 2019 is predicted to be 1366 t and unwanted catch to be 55 t . The SSB in 2020 is forecast to be 2013 t , which is well below $\mathrm{Blim}_{\text {lim }}$ (Table 5.18).

The forecast of landings in 2020 and SSB in 2021 in particular is sensitive to the recruitment assumptions. The assumption of GM recruitment in 2019 and 2020 contribute $41 \%$ and $28 \%$ respectively to the forecast SSB in 2021. (Figure 5.43).

### 3.5.1 Reference points

Both MSY and precautionary reference points were reconsidered at IBPCod.6a in February 2019 in accordance with ICES guidelines and are shown below (weights in tonnes). The final agreed reference points are given below. The estimate of $\mathrm{F}_{\text {msy }}$ is greater than previous due to i) exclusion of Beverton-Holt stock-recruitment relationship and ii) choice of yield to include catch above MCRS (estimated by assuming a historical discard rate).

|  | Advice 2015 | WKMSYREF4 | IBPCod.6a | Rationale (WKMSYREF4/IBPCod.6a) |
| :---: | :---: | :---: | :---: | :---: |
| $\mathrm{Bl}_{\text {lim }}$ | 14000 | 14000 | 14000 | Bloss from which the stock has increased (SSB in 1992 as estimated in 2015) |
| $\mathrm{B}_{\mathrm{pa}}$ | 22000 | 20000 | 20000 | $1.4 \times \mathrm{Bl}_{\text {lim }}$ |
| $F_{\text {lim }}$ | 0.8 | 0.82 | 0.77 | Based on simulation with segmented regression recruitment with Blim as the breakpoint |
| $\mathrm{F}_{\mathrm{pa}}$ | 0.6 | 0.59 | 0.55 | $\mathrm{F}_{\text {lim } / 1.4}$ |
| $\mathrm{F}_{\text {MSY }}$ | 0.19 | 0.167 | 0.29 | F that provides max yield (calculated from EqSim using Segmented regression \& Ricker stock-recruit relationship) |
| MSY $\mathrm{B}_{\text {trigger }}$ | 22000 | 20000 | 20000 | $\mathrm{B}_{\mathrm{pa}}$ |
| $\mathrm{F}_{\text {MSY }}$ upper |  | 0.254 | 0.41 | F at 95\% MSY (above $\mathrm{F}_{\text {MSY }}$ ) |
| $\mathrm{F}_{\text {MSY }}$ lower |  | 0.108 | 0.20 | F at 95\% MSY (below $\mathrm{F}_{\text {MSY }}$ ) |

### 3.5.2 Management plans

Technical measures applicable to the West of Scotland, including those associated with the cod recovery plan in force up to 2008 (Council Regulation No. 423/2004), the cod long-term management plan in force from 2009 (Council Regulation No. 1342/2008) were amended by Council Regulation No. 1243/2012. The management plan was further amended in 2016 by Council Regulation (EU) 2016/2094 to cover the transitional period in which preparations are ongoing towards multiannual plans for multispecies fisheries. In 2018 the cod management plan was discontinued. Cod in Division 6.a is not included in the multiannual plan for Western Waters (Council Regulation (EU) 2019/472).

### 3.6 Uncertainties and bias in assessment and forecast

Figure 5.44 shows a comparison between this year's and last year's assessments. Following the IBP (ICES, 2019), the assessment shows a substantial downward revision of mean F, since around 2010, compared to the assessment presented last year. The 2018 assessment estimated mean $F$ in 2017 to be 0.96 while in the current assessment this is now estimated as 0.60 . There is a similar upward revision in SSB: the 2018 assessment estimated SSB in 2017 to be 3413 tonnes while this year's assessment estimates it at 4196 t . These revisions are largely the result of changes to the TSA configuration agreed at the IBP in 2019.

An upward revision of recent estimates of SSB compared with that presented last year. The 2017 assessment estimated SSB in 2016 to be 2741 t while this year's assessment estimates it at 3630 t . Mean $F$ in that year is now estimated at 0.63 which is a significant downward revision compared to last year's assessment (0.94). The 2015 mean F has also been revised downwards.
The estimate of recruitment in 2018 is revised down from 2.826 million to 0.826 million (the lowest of the time-series) while recruitments from 2013-2015 have been revised upwards.

No forecast was conducted last year and therefore it is not possible to check the consistency of forecast SSB with current assessment estimates.

## Landings

Since the early 1990s the most significant problem with the assessment of this stock is with commercial data. Incorrect reporting of landings, species, quantity and management area, is known to have occurred. Scottish landings (from 2006) are adjusted to include estimates of misreporting (in an attempt to reduce bias in the assessment) and in 2016, area misreported landings account for around $55 \%$ of the total landings. The misreporting estimates are provided by Marine Scotland Compliance based on intelligence and consideration of VMS data. Estimates based on provisional analysis of VMS data linked to landings at a trip level (conducted at the 2015 interbenchmark (ICES, 2015)) gave somewhat higher estimates. In addition these misreported landings are unsampled and potentially have different age compositions to the rest of the Scottish demersal finfish fleet due to likely differences in discarding behaviour.

## Discards

On average (over the last five years), discarding accounts for almost $70 \%$ of the total catch. Although sampling levels have improved in recent years, discard estimates are still very uncertain (approximate $C V=50 \%$ for Scottish large mesh demersal fleet in 2017) contributing to uncertainty in the estimates of mean F.

## Biological factors

Assumptions on mean weight-at-length and mean maturity-at-age have remained unchanged for a long period. However, biological responses of cod in 6. a as a localised species to high exploitation and low population numbers are so far unknown to the working group.

The contribution of seal predation to total cod mortality is likely to be significant and this may impair the ability of the cod stock to recover but data is limited. Weight dependent natural mor-talities-at-age have been adopted to better take account of higher natural mortality at younger ages but it is not certain these values fully accommodate the possible large source of natural mortality from seals. Regular surveys giving estimates of consumption by seals would give greater confidence in natural mortality estimates. An assessment conducted by Cook et al. (2015)
suggests declining fishing mortality and that seal predation may be impairing the recovery of this stock.

## Stock structure

Stock structure is complex and at least two subpopulations are known to occur within this area. The survey distribution plots show that there is an almost complete absence of cod on the shelf in Division 6.a with the majority of the landings and stock concentrated in an area in the north of the region (around the 'windsock' closed area) bordering Division 4.a. It may be more appropriate to consider this component of the stock as part of the North Sea stock (or at least the northern component of this stock).

## Assessment method

The input data for this cod assessment are particularly uncertain (both survey indices and commercial data) and as a result, the data can be interpreted in different ways by different assessment methods. The assessment presented by Cook (2019) shows a stock which has recovered to levels consistent with those of the 1990s, while the ICES TSA assessment shows little sign of SSB recovery. The uncertainty estimates from the final TSA assessment are therefore unlikely to adequately reflect the true uncertainty in the estimates of stock biomass and fishing mortality for this stock.

### 3.6.1 Recommendation for next Benchmark

| problem | Solution | expertise necessary ${ }^{\mathbf{1}}$ | suggested time |
| :--- | :--- | :--- | :--- |
| Stock identity | Evaluate a possible merge between <br> North Sea and 6.a cod stocks. Or as an <br> alternative, split area 6.a in two areas <br> North and South. | Scientists from MSS and <br> MI | Next benchmark alt- <br> hough would need col- <br> laboration with <br> WGNSSK. |
| Misreporting of land- <br> ings; does not take ac- <br> count of fleet compo- <br> nents. | Further analysis of misreporting data <br> supplied by Scotland, potentially mak- <br> ing use of VMS data | Scientists from MSS | One year before the <br> benchmark as it is a <br> proceess that is time <br> consuming. |
| Fishery selectivity pat- <br> tern | Flat-topped \& dome-shaped selectiv- <br> ity pattern both plausible; modelling <br> the main fleets separately may help. <br> Will require a longer time-series of <br> fleet disaggregated data. | Scientists from MSS (and <br> MI) | A data call before the <br> data compilation work- <br> shop. |
| Discard proportions <br> poorly estimated | Alternative discard models: using an <br> age-based ogive or modelling all age <br> one as discards | Scientists from MSS; using <br> alternative assessment <br> models or bespoke modifi- <br> cations to TSA. | Ahead of benchmark |
| Assessment method | Application of alternative stock assess- <br> ment models and/or a multiple model <br> approach. | Scientists from MSS | In preparation for next <br> benchmark. |

[^3]
### 3.6.2 Management considerations

The fisheries for cod are fully under the landing obligation from 2019 onwards. In the past, they have been managed by a combination of landings limits, area closures and technical measures. The measures taken thus far have not recovered the stock. Although fishing mortality has reduced since 2005 (showing a proportionately similar reduction to the decline in reported effort since 2003), it remains well above Fmsy and showed an increase in 2019.

Cod are known to form aggregations, so it is still possible to find areas of high cod density at low stock abundance (as apparent in the Scottish Q1 survey in particular). This can lead to high caches in localized areas, generating high fishing mortality even with low fishing effort. The impact of this could potentially be reduced by the use of temporary spatial closures.

The fishing opportunities regulation explicitly made the stock a bycatch species from 2012 to 2018. Allowing landings up to $1.5 \%$ of the live weight of the total catch can cause a perverse incentive for vessels to increase catches of other species and does not inhibit the catch of cod. In fact, in recent years the landings of 6.a cod have increased.

Although the UK 'Buyers and Sellers' and Irish 'Sales Notes' legislation is considered to have reduced underreporting from 2006, discard data show increased discards at-ages one and two and a change in discard practices such that fish are discarded at older ages from this time onwards (i.e. the discards are now largely high-grading). With the full implementation of the landing obligation in 2019 for fisheries catching cod, a bycatch TAC of 1735 t has been set to allow mixed fisheries with a bycatch of cod to continue (Cf. ICES estimated catch of 1890 t in 2018 and 2363 t in 2017). It is not known how the fishery will respond to the increase in TAC. The forecast assumes that in 2019 and 2020, unwanted catch will be the catch below MCRS, which is approximated by the use of a historical discard proportion-at-age.
Estimates of misreporting (landings believed to be taken in Division 6.a and reported elsewhere) from Marine Scotland Compliance imply ICES landings estimates which are in excess of TACs during the mid-2000s. Area misreported landings account for over $60 \%$ of the total landings in 2018. Measures to reduce area misreporting should also be introduced.

Cod is taken in mixed demersal fisheries, and in Division 6.a is a bycatch species. To greatly reduce cod catch would likely result in having to greatly reduce harvesting of other stocks such as haddock, whiting and anglerfish. It is also important the bycatch from the Nephrops fleet is closely monitored (including discard observations). In 2018, large trawl gear vessels targeting finfish are responsible for $93 \%$ of cod catches in Division 6.a, the Nephrops fleet take approximately $4 \%$ and the remainder are taken by other gears, including longliners and gillnets.

A report by the Sea Mammal Research unit (Hammond and Harris, 2006) gives estimates of cod consumed by grey seals to the west of Scotland. Although highly uncertain, the estimates suggest predation mortality on cod is significant and this may impair the ability of the cod stock to recover, but data are limited (Cook et al., 2015).

### 3.6.3 Frequency of assessment

This stock has had zero catch advice for over ten years and therefore meets the first of the criteria for consideration for biennial assessment as outlined by WGCSE in 2016.

### 3.6.4 References

Cook, R. M. 2019. Stock collapse or stock recovery? Contrasting perceptions of a depleted cod stock. ICES Journal of Marine Science. doi:10.1093/icesjms/fsy190.

Cook, R. M., Holmes, S. J. and Fryer, R. J. 2015. Grey seal predation impairs recovery of an over-exploited fish stock. J. Applied Ecol., 52(4), 969-979.

EU. 2008. COUNCIL REGULATION (EC) No. 1342/2008 of 18 December 2008 establishing a long-term plan for cod stocks and the fisheries exploiting those stocks and repealing Regulation (EC) No. 423/2004. Official Journal of the European Union, L 348/21.
http://eur-lex.europa.eu/LexUriServ/LexUriServ.do?uri=OI:L:2008:348:0020:0033:EN:PDF.
EU. 2016. COUNCIL REGULATION (EU) 2016/2094 OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL of 23 November 2016 amending Council Regulation (EC) No 1342/2008 establishing a longterm plan for cod stocks and the fisheries exploiting those stocks.

Hammond, P. S., and Harris, R. N. 2006. Grey seal diet composition and prey consumption off western Scotland and Shetland. Final report to Scottish Executive Environment and Rural Affairs Department and Scottish Natural Heritage.

ICES. 2012. Report of the Benchmark Workshop on Western Waters Roundfish (WKROUND), 22-29 February 2012, Aberdeen, UK. ICES CM 2012/ACOM:49. 283 pp.
ICES. 2015. Report of the Inter-Benchmark Protocol of West of Scotland Roundfish (IBPWSRound), Feb-ruary-April 2015. ICES CM 2015/ACOM:37.

ICES. 2019. Inter-benchmark Workshop on West of Scotland Cod (6.a) (IBPCod6.a). ICES Scientific Reports. 1:13. 171 pp . http://doi.org/10.17895/ices.pub.4976.

Lorenzen K. 1996. The relationship between body weight and natural mortality in juvenile and adult fish: a comparison of natural ecosystems and aquaculture. Journal of Fish Biology 49, 627-647.

STECF. 2011. Scientific, Technical and Economic Committee for Fisheries. Evaluation of Fishing Effort Regimes Regarding Annexes IIA, IIB and IIC of TAC \& Quota Regulations, Celtic Sea and Bay of Biscay (STECF-11-13).

STECF. 2011. Scientific, Technical and Economic Committee for Fisheries. Evaluation of Fishing Effort Regimes (STECF-13-13).

Scientific, Technical and Economic Committee for Fisheries (STECF) - Evaluation of Fishing Effort Regimes in European Waters - Part 2 (STECF-14-20). 2014. Publications Office of the European Union, Luxembourg, EUR 27027 EN, JRC 93183, 844 pp.

Table 5.1. Cod in Division 6.a. ICES official catch statistics.

| Country | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 48 | 88 | 33 | 44 | 28 | - | 6 | - | 22 | 1 | 2 | + | 11 | 1 | + | + | 2 |
| Denmark | - | - | 4 | 1 | 3 | 2 | 2 | 3 | 2 | + | 4 | 2 | - | - | + | - | - |
| Faroe Islands | - | - | - | 11 | 26 | - | - | - | - | - | - | - | - | - | - | - | - |
| France | 7,411 | 5,096 | 5,044 | 7,669 | 3,640 | 2,220 | 2,503 | 1,957 | 3,047 | 2,488 | 2,533 | 2,253 | 956 | 714 | 842 | 236 | 391 |
| Germany | 66 | 53 | 12 | 25 | 281 | 586 | 60 | 5 | 94 | 100 | 18 | 63 | 5 | 6 | 8 | 6 | 4 |
| Greenland |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Ireland | 2,564 | 1,704 | 2,442 | 2,551 | 1,642 | 1,200 | 761 | 761 | 645 | 825 | 1,054 | 1,286 | 708 | 478 | 223 | 357 | 319 |
| Netherlands | - | - | - | - | - | - | - | - | - | - | - | - | 2 | 1 | - | - | - |
| Norway | 204 | 174 | 77 | 186 | 207 | 150 | 40 | 171 | 72 | 51 | 61 | 137 | 36 | 36 | 79 | 114 | 39 |
| Spain | 28 | - | - | - | 85 | - | - | - | - | - | 16 | + | 6 | 42 | 45 | 14 | 3 |
| UK (E. W. N.I.) | 260 | 160 | 444 | 230 | 278 | 230 | 511 | 577 | 524 | 419 | 450 | 457 | 779 | 474 | 381 | 280 | 138 |
| UK (Scotland) | 8,032 | 4,251 | 11,143 | 8,465 | 9,236 | 7,389 | 6,751 | 5,543 | 6,069 | 5,247 | 5,522 | 5,382 | 4,489 | 3,919 | 2,711 | 2,057 | 1544 |
| UK | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Total landings | 18,613 | 11,526 | 19,199 | 19,182 | 15,426 | 11,777 | 10,634 | 9,017 | 10,475 | 9,131 | 9,660 | 9,580 | 6,992 | 5,671 | 4,289 | 3,064 | 2440 |


| Country | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | + | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |
| Denmark | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 11.2 | 0.7 | 0.039 |
| Faroe Islands | - | - | 2 | 0 | 1 | 12 | 1 | - | - | - | - | - | - | - | - | - | 0.25 |
| France | 208 | 172 | 91 | 107 | 108 | 92 | 82 | 74 | 60 | 49 | 4 | 3 | 5 | 11.4 | 85.7 | 118.5 | 100.52 |
| Germany | + | + | - | - | 2 | 2 | 1 | - | - | - | - | - | - | - | - | - |  |
| Greenland |  |  | - | - |  |  |  | - | - | - | - | - | - | - | + | - |  |
| Ireland | 210 | 120 | 34 | 28 | 18 | 70 | 58 | 24 | 49 | 41 | 18 | 14 | 12 | 17.5 | 27.5 | 18.6 | 12.1505 |
| Netherlands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |
| Norway | 88 | 45 | 10 | 17 | 30 | 30 | 65 | 18 | 20 | 8 | 2 | 24 | 14 | 59 | 39.3 | 14 | 36.79 |
| Spain | 11 | 3 | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |
| UK (E. W. N.I.) | 195 | 79 | 46 | 25 | 14 | 21 | 6 | 14 | 4 | 3 | 2 | 1 | 9 |  |  |  |  |
| UK (Scotland) | 1519 | 879 | 413 | 243 | 318 | 260 | 232 | 104 | 115 | 107 | 135 | 130 | 121 |  |  |  |  |
| UK | - | - | - | - |  | - | - |  |  |  |  |  |  | 168 | 182.2 | 198.9 | 210.083 |
| Total landings | 2231 | 1298 | 596 | 420 | 491 | 487 | 445 | 234 | 248 | 208 | 161 | 172 | 161 | 255.9 | 346.7 | 350.7 | 359.8325 |

* Preliminary.
$+<1$ tonne

Table 5.2. Cod in Division 6.a. Landings (reported into 6a and area misreported), discards and catch (tonnes) estimates, as used by the WG. Values are totals for fish aged 1 to 7+.

| Year | Landings |  | Discards | Catch |
| :---: | :---: | :---: | :---: | :---: |
|  | reported | misreported |  |  |
| 1981 | 23865 |  | 303 | 24168 |
| 1982 | 21511 |  | 571 | 22082 |
| 1983 | 21305 |  | 197 | 21503 |
| 1984 | 21272 |  | 329 | 21601 |
| 1985 | 18607 |  | 963 | 19570 |
| 1986 | 11820 |  | 263 | 12083 |
| 1987 | 18971 |  | 2388 | 21358 |
| 1988 | 20413 |  | 368 | 20781 |
| 1989 | 17169 |  | 2076 | 19246 |
| 1990 | 12175 |  | 571 | 12746 |
| 1991 | 10927 |  | 622 | 11549 |
| 1992 | 9086 |  | 1779 | 10865 |
| 1993 | 10314 |  | 139 | 10453 |
| 1994 | 8928 |  | 661 | 9588 |
| 1995 | 9439 |  | 141 | 9580 |
| 1996 | 9427 |  | 63 | 9489 |
| 1997 | 7034 |  | 499 | 7533 |
| 1998 | 5714 |  | 538 | 6252 |
| 1999 | 4201 |  | 69 | 4270 |
| 2000 | 2977 |  | 821 | 3798 |
| 2001 | 2347 |  | 92 | 2439 |
| 2002 | 2243 |  | 480 | 2722 |
| 2003 | 1241 |  | 34 | 1275 |
| 2004 | 540 |  | 72 | 612 |
| 2005 | 511 |  | 41 | 552 |
| 2006 | 463 | 26 | 465 | 954 |
| 2007 | 524 | 70 | 1880 | 2474 |
| 2008 | 454 | 228 | 695 | 1377 |
| 2009 | 222 | 186 | 945 | 1353 |
| 2010 | 239 | 320 | 785 | 1344 |
| 2011 | 170 | 284 | 1670 | 2124 |
| 2012 | 174 | 292 | 1166 | 1632 |
| 2013 | 176 | 123 | 1202 | 1501 |
| 2014 | 152 | 205 | 1311 | 1668 |
| 2015 | 308 | 461 | 983 | 1752 |
| 2016 | 394 | 498 | 852 | 1745 |
| 2017 | 365 | 429 | 1569 | 2363 |
| 2018 | 388 | 741 | 760 | 1890 |

Table 5.3. Cod in Division 6.a. Survey data made available to the WG. Data used in update assessment are highlighted in bold. For the Scottish surveys, numbers are standardised to catch rate per ten hours. For the Irish surveys, effort is given as minutes towed and numbers are in units.

ScoGFS- WIBTS- Q1: Scottish west coast groundfish survey

| 1985 | 2010 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0 | 0.25 |  |  |  |  |  |
| 1 | 7 |  |  |  |  |  |  |  |
| 10 | 1.5 | 23.7 | 8.6 | 13.6 | 3.9 | 2.5 | 1.2 | 1985 |
| 10 | 1.5 | 6.9 | 26.8 | 5.6 | 7.3 | 2.5 | 1.9 | 1986 |
| 10 | 57.4 | 16.2 | 15.3 | 22.8 | 3.0 | 2.8 | 0.0 | 1987 |
| 10 | 0.0 | 64.9 | 14.2 | 3.4 | 2.1 | 0.7 | 0.2 | 1988 |
| 10 | 4.5 | 7.2 | 45.1 | 8.6 | 1.9 | 0.5 | 0.8 | 1989 |
| 10 | 2.0 | 24.6 | 4.1 | 14.7 | 4.2 | 1.6 | 0.8 | 1990 |
| 10 | 4.8 | 5.4 | 17.4 | 5.2 | 13.4 | 2.8 | 0.5 | 1991 |
| 10 | 7.3 | 11.5 | 5.4 | 7.6 | 3.4 | 2.3 | 0.5 | 1992 |
| 10 | 1.7 | 38.2 | 12.7 | 1.7 | 1.4 | 1.1 | 0.0 | 1993 |
| 10 | 13.6 | 14.7 | 25.1 | 5.8 | 1.0 | 0.0 | 0.0 | 1994 |
| 10 | 6.4 | 23.8 | 14.0 | 16.5 | 1.2 | 1.9 | 0.7 | 1995 |
| 10 | 2.8 | 20.9 | 24.1 | 4.1 | 2.8 | 1.3 | 0.0 | 1996 |
| 10 | 11.1 | 7.7 | 11.6 | 7.9 | 4.2 | 4.7 | 1.0 | 1997 |
| 10 | 2.8 | 30.9 | 5.3 | 8.7 | 3.7 | 0.6 | 2.0 | 1998 |
| 10 | 1.5 | 8.2 | 8.2 | 1.4 | 3.2 | 0.5 | 0.5 | 1999 |
| 10 | 13.3 | 5.4 | 6.9 | 1.3 | 0.0 | 0.4 | 0.0 | 2000 |
| 10 | 2.7 | 18.4 | 5.7 | 13.2 | 19.5 | 1.1 | 1.6 | 2001 |
| 10 | 5.3 | 4.3 | 10.6 | 2.6 | 0.5 | 3.0 | 0.0 | 2002 |
| 10 | 2.7 | 16.7 | 2.0 | 4.7 | 1.8 | 0.7 | 0.4 | 2003 |
| 10 | 5.7 | 3.0 | 5.6 | 2.3 | 1.7 | 0.0 | 0.0 | 2004 |
| 10 | 1.3 | 1.5 | 1.2 | 0 | 0 | 0.4 | 0 | 2005 |
| 10 | 2.2 | 1.9 | 1.1 | 0.3 | 0 | 0 | 0.3 | 2006 |
| 10 | 2.1 | 18.8 | 3.4 | 1.2 | 0 | 0.6 | 0 | 2007 |
| 10 | 0.8 | 2.1 | 44.2 | 6.3 | 0.8 | 0 | 0 | 2008 |
| 10 | 1.8 | 2.6 | 2.3 | 0.4 | 0 | 0 | 0 | 2009 |
| 10 | 4.6 | 16.2 | 3.7 | 1.0 | 0.7 | 0 | 0 | 2010 |

Table 5.3. Continued. Cod in Division 6.a. Survey data made available to the WG. For the Scottish surveys, numbers are standardised to catch rate per ten hours. For the Irish surveys, effort is given as minutes towed and numbers are in units.

UK-SCOWCGFS-Q1 (index)

| 2011 | 2019 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0 | 0.25 |  |  |  |  |  |
| 1 | 7 |  |  |  |  |  |  |  |
| 10 | 0.52 | 32.95 | 21.07 | 0.93 | 0.98 | 0.74 | 0.00 | 2011 |
| 10 | 13.99 | 27.30 | 22.72 | 4.58 | 3.50 | 2.20 | 4.20 | 2012 |
| 10 | 20.03 | 40.26 | 26.38 | 36.95 | 7.76 | 0.30 | 0.00 | 2013 |
| 10 | 11.40 | 41.73 | 13.44 | 5.12 | 4.31 | 0.75 | 0.00 | 2014 |
| 10 | 8.16 | 36.40 | 70.70 | 37.74 | 23.25 | 13.00 | 2.47 | 2015 |
| 10 | 4.73 | 56.07 | 65.41 | 44.56 | 5.67 | 2.36 | 2.29 | 2016 |
| 10 | 2.92 | 33.49 | 50.58 | 49.58 | 156.64 | 10.71 | 24.89 | 2017 |
| 10 | 1.728 | 20.375 | 7.199 | 19.765 | 9.98 | 2.261 | 1.092 | 2018 |
| 10 | 9.924 | 4.173 | 6.888 | 2.031 | 3.181 | 0.318 | 0.318 | 2019 |

UK-SCOWCGFS-Q1 (variance)

| 2011 | 2019 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0 | 0.25 |  |  |  |  |  |
| 1 | 7 |  |  |  |  |  |  |  |
| 10 | 0.09 | 78.37 | 24.06 | 0.22 | 0.49 | 0.30 | 0.00 | 2011 |
| 10 | 44.18 | 120.08 | 33.73 | 2.31 | 8.34 | 4.83 | 13.02 | 2012 |
| 10 | 118.35 | 151.04 | 136.89 | 240.05 | 6.47 | 0.09 | 0.00 | 2013 |
| 10 | 20.17 | 383.27 | 12.23 | 3.04 | 5.47 | 0.28 | 0.00 | 2014 |
| 10 | 14.35 | 112.82 | 1264.73 | 602.27 | 289.82 | 98.91 | 5.48 | 2015 |
| 10 | 1.81 | 214.42 | 607.48 | 319.21 | 5.02 | 1.60 | 1.85 | 2016 |
| 10 | 1.43 | 155.67 | 498.57 | 1061.90 | 20475.95 | 84.79 | 287.62 | 2017 |
| 10 | 1 | 24.03 | 2.21 | 20.09 | 7.46 | 0.5 | 0.25 | 2018 |
| 10 | 6.79 | 2.03 | 6.12 | 0.6 | 1.98 | 0.1 | 0 | 2019 |

Table 5.3. Continued. Cod in Division 6.a. Survey data made available to the WG. For the Scottish surveys, numbers are standardised to catch rate per ten hours. For the Irish surveys, effort is given as minutes towed and numbers are in units.

| IreGFS | Irish groundfish survey |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| 1993 | 2002 |  |  |  |
| 1 | 1 | 0.75 | 0.79 |  |
| 0 | 3 |  |  |  |
| 1849 | 0.0 | 312.0 | 49.0 | 13.0 |
| 1610 | 20.0 | 999.0 | 56.0 | 13.0 |
| 1826 | 78.0 | 169.0 | 142.0 | 69.0 |
| 1765 | 0.0 | 214.0 | 89.0 | 18.0 |
| 1581 | 6.0 | 565.0 | 31.0 | 10.0 |
| 1639 | 0.0 | 83.0 | 53.0 | 6.0 |
| 1564 | 0.0 | 24.0 | 14.0 | 3.0 |
| 1556 | 0.0 | 124.0 | 4.0 | 1.0 |
| 755 | 3.0 | 82.0 | 28.0 | 2.0 |
| 798 | 0.0 | 50.6 | 2.2 | 1.2 |

ScoGFS-WIBTS-Q4: Quarter 4 Scottish ground fish survey

| $\begin{gathered} 1996 \\ \hline 1 \\ \hline \end{gathered}$ | 2010 |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0.75 | 1.00 |  |  |  |  |  |  |  |
| 0 | 8 |  |  |  |  |  |  |  |  |  |
| 10 | 0 | 0.7 | 14.4 | 5 | 3 | 1.1 | 0.5 | 0 | 0 | 1996 |
| 10 | 1 | 10.9 | 2.4 | 1.4 | 1.4 | 1.4 | 0.2 | 0 | 0 | 1997 |
| 10 | $+$ | 14.8 | 9.7 | 1.1 | 0 | 0 | 0 | 0 | 0 | 1998 |
| 10 | 2 | 4 | 6 | 9.2 | 0.5 | 0 | 0 | 0 | 0 | 1999 |
| 10 | 0 | 15.8 | 2.6 | 0.4 | 0.4 | 0 | 0 | 0 | 0 | 2000 |
| 10 | 1 | 1.7 | 7.3 | 1.7 | 0.3 | 0 | 0 | 0 | 0 | 2001 |
| 10 | 1 | 10.4 | 2.8 | 6.8 | 0.6 | 0 | 0 | 0 | 0 | 2002 |
| 10 | 1 | 1.5 | 11.3 | 2.9 | 0.6 | 0 | 0 | 0 | 0 | 2003 |
| 10 | 0 | 5.1 | 3.8 | 1.4 | 0 | 0.7 | 0 | 0 | 0 | 2004 |
| 10 | + | 2.1 | 3 | 0 | 0.6 | 0.3 | 0 | 0 | 0 | 2005 |
| 10 | 0 | 16.9 | 5.9 | 1.4 | 0.7 | 0 | 0 | 0 | 0 | 2006 |
| 10 | 0 | 12 | 20 | 1.3 | 0.5 | 0 | 0.3 | 0 | 0 | 2007 |
| 10 | 2 | 7.7 | 5 | 7 | 1 | 0 | 0 | 0 | 0 | 2008 |
| 10 | 2 | 14.2 | 3.8 | 1.2 | 1.2 | 0.3 | 0 | 0 | 0 | 2009 |
| 10 | na | na | na | na | na | na | na | na | na | 2010 |

Table 5.3. Cont. Cod in Division 6.a. Survey data made available to the WG. For the Scottish surveys, numbers are standardised to catch rate per ten hours. For the Irish surveys, effort is given as minutes towed and numbers are in units.

UK-SCOWCGFS-Q4 (index)

| 2011 | 2018 | 0.75 | 1.0 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 |  |  |  |  |  |  |  |  |  |
| 0 | 8 |  |  |  |  |  |  |  |  |  |
| 10 | 0.60 | 9.71 | 31.54 | 10.88 | 0.93 | 0.00 | 0.00 | 2011 |  |  |
| 10 | 0.75 | 19.78 | 7.12 | 15.43 | 13.60 | 1.02 | 0.68 | 0.34 | 0.00 | 2012 |
| 10 | 1.67 | 23.65 | 28.06 | 15.63 | 5.57 | 6.63 | 1.37 | 0.00 | 0.00 | 2014 |
| 10 | 3.64 | 28.17 | 52.53 | 34.22 | 10.58 | 4.24 | 5.27 | 1.18 | 0.59 | 2015 |
| 10 | 0.374 | 6.162 | 34.941 | 45.443 | 118.92 | 14.893 | 5.773 | 3.176 | 0 | 2016 |
| 10 | 2.127 | 10.024 | 6.221 | 24.427 | 10.881 | 8.538 | 0.767 | 0.511 | 0 | 2017 |
| 10 | 0 | 4.569 | 15.945 | 4.809 | 39.902 | 29.022 | 10.887 | 0.829 | 0 | 2018 |
|  |  |  |  |  |  |  |  |  |  |  |

UK-SCOWCGFS-Q4 (variance)


Table 5.3. Continued. Cod in Division 6.a. Survey data made available to the WG. For the Scottish surveys, numbers are standardised to catch rate per ten hours. For the Irish surveys, effort is given as minutes towed and numbers are in units.

IRGFS-WIBTS-Q4 Irish West Coast groundfish.

| 2003 | 2018 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0.79 | 0.92 |  |  |  |
| 0 | 4 |  |  |  |  |  |
| 1127 | 0 | 10 | 11 | 0 | 0 | 2003 |
| 1200 | 0 | 24 | 10 | 1 | 0 | 2004 |
| 960 | 63 | 13 | 7 | 0 | 2 | 2005 |
| 1510 | 0 | 95 | 12 | 0 | 0 | 2006 |
| 1173 | 0 | 161 | 12 | 0 | 1 | 2007 |
| 1135 | 0 | 23 | 24 | 4 | 0 | 2008 |
| 1378 | 1 | 75 | 4 | 5 | 0 | 2009 |
| 1291 | 0 | 70 | 31 | 4 | 3 | 2010 |
| 1287 | 1 | 26 | 26 | 4 | 0 | 2011 |
| 1230 | 0 | 74 | 7 | 3 | 0 | 2012 |
| 1295 | 0 | 92 | 11 | 0 | 0 | 2013 |
| 1200 | 0 | 113 | 20 | 2 | 0 | 2014 |
| 1213 | 0 | 15 | 11 | 3 | 0 | 2015 |
| 962 | 0 | 27 | 23 | 2 | 0 | 2016 |
| 1196 | 0 | 2 | 17 | 7 | 2 | 2017 |
| 966 | 1 | 21 | 3 | 0 | 1 | 2018 |

Table 5.4. Cod in Division 6.a. Landings-at-age (thousands). Values for 2006 onwards include an adjustment for area misreporting.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 723 | 1761 | 999 | 695 | 286 | 97 | 75 |
| 1979 | 929 | 1612 | 2125 | 682 | 342 | 134 | 69 |
| 1980 | 1195 | 3294 | 2001 | 796 | 191 | 77 | 37 |
| 1981 | 461 | 7016 | 3220 | 904 | 182 | 29 | 20 |
| 1982 | 1827 | 1673 | 3206 | 1189 | 367 | 111 | 33 |
| 1983 | 2335 | 4515 | 1118 | 1400 | 468 | 148 | 60 |
| 1984 | 2143 | 2360 | 2564 | 448 | 555 | 185 | 59 |
| 1985 | 1355 | 5069 | 1269 | 1091 | 140 | 167 | 79 |
| 1986 | 792 | 1486 | 2055 | 411 | 191 | 40 | 30 |
| 1987 | 7873 | 4837 | 988 | 905 | 137 | 56 | 26 |
| 1988 | 1008 | 8336 | 2193 | 278 | 210 | 39 | 20 |
| 1989 | 2017 | 1082 | 3858 | 709 | 113 | 69 | 33 |
| 1990 | 513 | 4024 | 432 | 924 | 170 | 23 | 11 |
| 1991 | 1518 | 1728 | 1805 | 188 | 266 | 70 | 23 |
| 1992 | 1407 | 1868 | 575 | 720 | 69 | 58 | 24 |
| 1993 | 328 | 3596 | 1050 | 131 | 183 | 24 | 36 |
| 1994 | 942 | 1207 | 1545 | 280 | 56 | 51 | 20 |
| 1995 | 753 | 2750 | 700 | 630 | 70 | 15 | 11 |
| 1996 | 341 | 2331 | 1210 | 247 | 204 | 31 | 13 |
| 1997 | 1414 | 1067 | 989 | 281 | 66 | 62 | 7 |
| 1998 | 310 | 3318 | 293 | 174 | 57 | 16 | 9 |
| 1999 | 132 | 884 | 1047 | 64 | 48 | 24 | 9 |
| 2000 | 765 | 532 | 211 | 231 | 15 | 12 | 13 |
| 2001 | 96 | 1241 | 155 | 63 | 52 | 3 | 4 |
| 2002 | 337 | 340 | 522 | 41 | 13 | 14 | 4 |
| 2003 | 62 | 516 | 85 | 107 | 6 | 2 | 1 |
| 2004 | 44 | 92 | 85 | 11 | 26 | 2 | 1 |
| 2005 | 31 | 121 | 43 | 37 | 7 | 6 | 0 |
| 2006 | 18 | 96 | 76 | 22 | 13 | 2 | 1 |
| 2007 | 6 | 187 | 70 | 37 | 3 | 4 | 3 |
| 2008 | 0 | 34 | 130 | 25 | 16 | 1 | 3 |
| 2009 | 2 | 12 | 11 | 59 | 8 | 2 | 0 |
| 2010 | 0 | 43 | 61 | 38 | 32 | 1 | 0 |
| 2011 | 0 | 11 | 40 | 34 | 12 | 13 | 2 |
| 2012 | 3 | 1 | 41 | 51 | 5 | 4 | 5 |
| 2013 | 0 | 8 | 9 | 43 | 10 | 2 | 1 |
| 2014 | 0 | 3 | 66 | 31 | 23 | 2 | 0 |
| 2015 | 0 | 53 | 55 | 41 | 29 | 27 | 1 |
| 2016 | 2 | 33 | 112 | 69 | 22 | 11 | 14 |
| 2017 | 1 | 22 | 62 | 54 | 52 | 15 | 3 |
| 2018 | 2 | $24$ | $45$ | $87$ | 55 | 34 | 11 |

Table 5.5. Cod in Division 6.a. Mean weight-at-age in landings (kg).

|  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 5.6. Cod in Division 6.a. Discard numbers-at-age (thousands).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 412 | 26 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 16 | 81 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 1171 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 54 | 907 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 1808 | 8 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 843 | 25 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 1088 | 11 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 5188 | 114 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 970 | 14 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 14358 | 12 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 231 | 1059 | 2 | 0 | 0 | 0 | 0 |
| 1989 | 6243 | 6 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 4181 | 41 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 2518 | 14 | 2 | 0 | 0 | 0 | 0 |
| 1992 | 7385 | 143 | 3 | 0 | 0 | 0 | 0 |
| 1993 | 279 | 84 | 1 | 0 | 0 | 0 | 0 |
| 1994 | 2743 | 6 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 625 | 56 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 191 | 50 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 1521 | 34 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 790 | 972 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 230 | 5 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 2882 | 33 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 176 | 115 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 1051 | 199 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 69 | 26 | 1 | 0 | 0 | 0 | 0 |
| 2004 | 232 | 21 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 108 | 20 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 1210 | 47 | 24 | 2 | 3 | 1 | 0 |
| 2007 | 566 | 1489 | 50 | 38 | 3 | 3 | 0 |
| 2008 | 68 | 102 | 281 | 1 | 0 | 0 | 0 |
| 2009 | 605 | 150 | 109 | 94 | 0 | 5 | 0 |
| 2010 | 352 | 392 | 65 | 7 | 3 | 0 | 0 |
| 2011 | 316 | 281 | 535 | 42 | 0 | 2 | 0 |
| 2012 | 374 | 93 | 383 | 50 | 0 | 0 | 0 |
| 2013 | 2030 | 321 | 131 | 103 | 15 | 0 | 2 |
| 2014 | 705 | 316 | 255 | 51 | 19 | 1 | 0 |
| 2015 | 161 | 307 | 217 | 25 | 6 | 1 | 0 |
| 2016 | 1008 | 209 | 95 | 46 | 6 | 0 | 0 |
| 2017 | 168 | 294 | 376 | 92 | 17 | 2 | 0 |
| 2018 | 93 | 146 | 60 | 87 | 17 | 7 | 2 |

Table 5.7. Cod in Division 6.a. Mean weight-at-age in discards (kg).

|  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 5.8. Cod in Division 6.a. Total catch-at-age (thousands).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 1135 | 1787 | 999 | 695 | 286 | 97 | 75 |
| 1979 | 945 | 1693 | 2125 | 682 | 342 | 134 | 69 |
| 1980 | 2366 | 3294 | 2001 | 796 | 191 | 77 | 37 |
| 1981 | 515 | 7923 | 3220 | 904 | 182 | 29 | 20 |
| 1982 | 3635 | 1681 | 3206 | 1189 | 367 | 111 | 33 |
| 1983 | 3178 | 4540 | 1118 | 1400 | 468 | 148 | 60 |
| 1984 | 3231 | 2371 | 2564 | 448 | 555 | 185 | 59 |
| 1985 | 6543 | 5183 | 1269 | 1091 | 140 | 167 | 79 |
| 1986 | 1762 | 1500 | 2055 | 411 | 191 | 40 | 30 |
| 1987 | 22231 | 4849 | 988 | 905 | 137 | 56 | 26 |
| 1988 | 1239 | 9395 | 2195 | 278 | 210 | 39 | 20 |
| 1989 | 8260 | 1088 | 3858 | 709 | 113 | 69 | 33 |
| 1990 | 4694 | 4065 | 432 | 924 | 170 | 23 | 11 |
| 1991 | 4036 | 1742 | 1807 | 188 | 266 | 70 | 23 |
| 1992 | 8792 | 2011 | 578 | 720 | 69 | 58 | 24 |
| 1993 | 607 | 3680 | 1051 | 131 | 183 | 24 | 36 |
| 1994 | 3685 | 1213 | 1545 | 280 | 56 | 51 | 20 |
| 1995 | 1378 | 2806 | 700 | 630 | 70 | 15 | 11 |
| 1996 | 532 | 2381 | 1210 | 247 | 204 | 31 | 13 |
| 1997 | 2935 | 1101 | 989 | 281 | 66 | 62 | 7 |
| 1998 | 1100 | 4290 | 293 | 174 | 57 | 16 | 9 |
| 1999 | 362 | 889 | 1047 | 64 | 48 | 24 | 9 |
| 2000 | 3647 | 565 | 211 | 231 | 15 | 12 | 13 |
| 2001 | 272 | 1356 | 155 | 63 | 52 | 3 | 4 |
| 2002 | 1388 | 539 | 522 | 41 | 13 | 14 | 4 |
| 2003 | 131 | 542 | 86 | 107 | 6 | 2 | 1 |
| 2004 | 276 | 113 | 85 | 11 | 26 | 2 | 1 |
| 2005 | 139 | 141 | 43 | 37 | 7 | 6 | 0 |
| 2006 | 1228 | 143 | 100 | 24 | 16 | 2 | 1 |
| 2007 | 572 | 1677 | 120 | 75 | 6 | 7 | 3 |
| 2008 | 68 | 136 | 411 | 26 | 16 | 1 | 3 |
| 2009 | 607 | 162 | 120 | 154 | 8 | 7 | 0 |
| 2010 | 352 | 436 | 126 | 45 | 35 | 1 | 0 |
| 2011 | 316 | 292 | 574 | 77 | 12 | 15 | 2 |
| 2012 | 377 | 95 | 424 | 102 | 5 | 4 | 5 |
| 2013 | 2030 | 329 | 139 | 146 | 25 | 2 | 3 |
| 2014 | 705 | 320 | 322 | 81 | 42 | 3 | 0 |
| 2015 | 161 | 360 | 272 | 66 | 35 | 27 | 1 |
| 2016 | 1010 | 242 | 208 | 115 | 29 | 11 | 14 |
| 2017 | 169 | 316 | 437 | 145 | 69 | 17 | 3 |
| 2018 | 96 | 170 | 105 | 174 | 72 | 41 | 13 |

Table 5.9. Cod in Division 6.a. Mean weight-at-age (kg) in total catch.

|  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 5.10. Cod in Division 6.a. TSA parameter settings for the assessment run.

| Parameter | Setting | Justification |
| :--- | :--- | :--- |
| Age of full selection. | a $_{\mathrm{m}}=6$ | To allow flexibility when estimating fishery <br> selectivity. |
| Survey catchability <br> model | WIBTS.Q1 \& WIBTS.Q4: no transitory or <br> persistent changes |  |
|  | SCO.Q1, SCO.Q4 \& IRGFS.Q4: transitory changes <br> estimates | Allows for survey year effects |


| Parameter | Setting |
| :--- | :--- |
| Large year classes. | The 1986 year class was large, and recruitment at-age 1 in 1987 is not well modelled by the <br> Ricker recruitment model. Instead, $N(1,1987)$ is taken to be normally distributed with mean <br> $5 \eta 1 S$ exp $(-\eta 2 S)$. The factor of 5 was chosen by comparing maximum recruitment to median <br> recruitment from 1966-1996 for 6.a cod, haddock, and whiting in turn using previous XSA runs. <br> The coefficient of variation is again assumed to be constant. |
|  | The |
|  |  |

Table 5.11. Cod in Division 6.a. Comparison of TSA parameter estimates from recent assessments.

| Parameter | Notation | Description | $\begin{aligned} & 2018 \\ & \text { WG } \end{aligned}$ | 2019 IBP | 2019 WG |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Initial fishing mortality | $F(1,1981)$ | Fishing mortality-at-age $a$ in year $y$ | 0.323 | 0.2507 | 0.264 |
|  | $F(2,1981)$ |  | 0.668 | 0.5698 | 0.597 |
|  | $F\left(\mathrm{a}_{\mathrm{m}}, 1981\right)$ |  | 1.032 | 0.7805 | 0.753 |
| Fishing mortality standard deviations | $\sigma_{F}$ | Transitory changes in overall fishing mortality | 0.152 | 0.1304 | 0.168 |
|  | $\sigma_{u}$ | Persistent changes in selection (age effect in F) | 0.009 | 0.0299 | 0.031 |
|  | $\sigma_{v}$ | Transitory changes in the year effect in fishing mortality | 0.178 | 0.0398 | 0.041 |
|  | $\sigma_{Y}$ | Persistent changes in the year effect in fishing mortality | 0.000 | 0.0989 | 0.085 |
| Measurement CVs | $\mathrm{CV}_{\text {landings }}$ | CV of landings-at-age data | 0.125 | 0.0881 | 0.067 |
|  | $\mathrm{CV}_{\text {discards }}$ | CV of discards-at-age data | 0.445 | 0.5776 | 0.558 |
| Recruitment | $\eta_{1}$ | Ricker parameter (slope at the origin) | 1.282 | 1.1145 | 1.029 |
|  | $\eta_{2}$ | Ricker parameter (curve dome occurs at $1 / \eta_{2}$ ) | 0.024 | 0.0203 | 0.017 |
|  | $c V_{\text {rec }}$ | Coefficient of variation of recruitment data | 0.407 | 0.4213 | 0.459 |
| Discards | $\sigma_{\text {logit }}$ | Transitory trends in discarding | 0.788 | 0.7766 | 0.731 |
|  | $\sigma_{\text {persistent }}$ | Persistent trends in discarding | 0.296 | 0.2428 | 0.248 |
|  | Step fn age 1 | Amount by which discards increase in 2006 | 4.058 | 5.9209 | 5.211 |
|  | Step fn age 2 |  | 5.895 | 6.2889 | 5.789 |
|  | Step fn age 3 |  | 0.985 | 1.1065 | 0.913 |
|  | Step fn age 4 |  | -0.436 | -0.1026 | -0.173 |


| Parameter | Notation | Description | $\begin{aligned} & 2018 \\ & \text { WG } \end{aligned}$ | 2019 IBP | 2019 WG |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Survey selectivities WCIBTS.Q1 | $\Phi(1)$ | Survey selectivity-at-age a | 0.561 | 0.4805 | 0.509 |
|  | $\Phi(2)$ |  | 2.897 | 2.8158 | 2.976 |
|  | $\Phi(3)$ |  | 6.950 | 6.4213 | 6.745 |
|  | $\Phi(4)$ |  | 10.666 | 9.9981 | 10.176 |
|  | $\Phi(5)$ |  | 15.379 | 12.9927 | 12.981 |
|  | $\Phi(6)$ |  | 20.789 | 15.1818 | 14.852 |
| Survey CVs | $\sigma_{\text {survey }}$ | CV parameter controlling gamma type dispersion | 0.258 | 0.0393 | 0.014 |
|  | $\eta_{\text {survey }}$ | CV parameter controlling poisson type dispersion | 1.142 | 1.5815 | 1.649 |
| Survey catchability standard deviations | $\sigma_{\Omega}$ | Transitory changes in survey catchability | NA | NA | NA |
|  | $\sigma_{\beta}$ | Persistent changes in survey catchability | NA | NA | NA |
| Survey selectivities UK-SCO.Q1 | $\Phi(1)$ | Survey selectivity-at-age a | 0.841 | 0.6911 | 1.393 |
|  | $\Phi(2)$ |  | 20.677 | 23.4037 | 27.540 |
|  | $\Phi(3)$ |  | 40.604 | 37.077 | 39.420 |
|  | $\Phi(4)$ |  | 49.005 | 48.9306 | 48.387 |
|  | $\Phi(5)$ |  | 84.270 | 71.0896 | 71.739 |
|  | $\Phi(6)$ |  | 63.453 | 48.8489 | 35.339 |
| Survey catchability standard deviations | $\sigma_{\Omega}$ | Transitory changes in survey catchability | 0.388 | 0.3794 | 0.302 |
|  | $\sigma_{\beta}$ | Persistent changes in survey catchability | NA | NA | NA |
| Survey selectivities WCIBTS.Q4 | $\Phi(1)$ | Survey selectivity-at-age a | NA | 3.1029 | 3.410 |
|  | $\Phi(2)$ |  | NA | 6.2709 | 7.131 |
|  | $\Phi(3)$ |  | NA | 5.1223 | 6.019 |
|  | $\Phi(4)$ |  | NA | 1.9957 | 2.348 |
| Survey CVs | $\sigma_{\text {survey }}$ | CV parameter controlling gamma type dispersion | NA | 0.0498 | 0.044 |
|  | $\eta_{\text {survey }}$ | CV parameter controlling poisson type dispersion | NA | 2.643 | 2.497 |


| Parameter | Notation | Description | 2018 | WG | 2019 IBP |
| :--- | :--- | :--- | :--- | :--- | :--- | 2019 WG

Table 5.12. Cod in Division 6.a. TSA population numbers-at-age (thousands).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 11028 | 19248 | 6959 | 1965 | 433 | 68 | 48 |
| 1982 | 26437 | 5488 | 7166 | 2456 | 739 | 183 | 48 |
| 1983 | 15317 | 12839 | 2398 | 2647 | 885 | 268 | 87 |
| 1984 | 27103 | 6437 | 5142 | 806 | 847 | 295 | 110 |
| 1985 | 13406 | 12755 | 2482 | 1607 | 235 | 203 | 115 |
| 1986 | 23663 | 4839 | 4328 | 778 | 347 | 72 | 80 |
| 1987 | 57984 | 10899 | 1902 | 1497 | 252 | 111 | 52 |
| 1988 | 6830 | 19733 | 4131 | 567 | 368 | 80 | 52 |
| 1989 | 24723 | 2795 | 6630 | 1217 | 191 | 112 | 47 |
| 1990 | 8515 | 10036 | 1004 | 1631 | 340 | 53 | 42 |
| 1991 | 13386 | 3497 | 3633 | 363 | 479 | 118 | 35 |
| 1992 | 22922 | 5508 | 1127 | 1250 | 133 | 150 | 48 |
| 1993 | 9087 | 10350 | 2110 | 337 | 347 | 45 | 72 |
| 1994 | 18653 | 4156 | 4050 | 611 | 118 | 102 | 39 |
| 1995 | 14871 | 8100 | 1723 | 1441 | 179 | 36 | 40 |
| 1996 | 6364 | 6757 | 2696 | 545 | 439 | 59 | 26 |
| 1997 | 23205 | 2879 | 2111 | 708 | 155 | 125 | 21 |
| 1998 | 6608 | 10072 | 741 | 484 | 199 | 44 | 37 |
| 1999 | 4501 | 2691 | 2746 | 175 | 132 | 66 | 24 |
| 2000 | 16469 | 1937 | 728 | 666 | 46 | 37 | 28 |
| 2001 | 3566 | 6129 | 556 | 199 | 171 | 12 | 17 |
| 2002 | 7436 | 1510 | 1886 | 141 | 44 | 43 | 9 |
| 2003 | 1978 | 2714 | 425 | 490 | 34 | 9 | 11 |
| 2004 | 2947 | 766 | 652 | 108 | 123 | 10 | 5 |
| 2005 | 1660 | 1168 | 219 | 188 | 34 | 28 | 3 |
| 2006 | 5291 | 661 | 395 | 48 | 43 | 7 | 8 |
| 2007 | 2088 | 2332 | 213 | 109 | 9 | 10 | 4 |
| 2008 | 1424 | 825 | 775 | 53 | 28 | 2 | 3 |
| 2009 | 3681 | 639 | 302 | 237 | 13 | 9 | 1 |
| 2010 | 3642 | 1643 | 244 | 88 | 68 | 3 | 3 |
| 2011 | 1586 | 1727 | 665 | 69 | 23 | 23 | 2 |
| 2012 | 2407 | 716 | 737 | 164 | 12 | 7 | 7 |
| 2013 | 4986 | 1066 | 323 | 265 | 48 | 4 | 5 |
| 2014 | 3592 | 1925 | 467 | 138 | 100 | 16 | 3 |
| 2015 | 3337 | 1685 | 906 | 178 | 62 | 44 | 8 |
| 2016 | 1282 | 1655 | 773 | 379 | 68 | 21 | 19 |
| 2017 | 1579 | 592 | 813 | 318 | 161 | 27 | 15 |
| 2018 | 826 | 764 | 240 | 339 | 128 | 65 | 15 |
| 2019* | 6404 | 381 | 345 | 89 | 121 | 41 | 27 |

*2019 values are TSA-derived projections of population numbers (smoothed).

Table 5.13. Cod in Division 6.a. Standard errors on TSA population numbers-at-age (thousands).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 1263 | 875 | 319 | 102 | 34 | 7 | 7 |
| 1982 | 1131 | 191 | 334 | 148 | 58 | 24 | 5 |
| 1983 | 1198 | 456 | 87 | 153 | 82 | 38 | 12 |
| 1984 | 1239 | 359 | 216 | 43 | 81 | 53 | 25 |
| 1985 | 1278 | 447 | 133 | 96 | 23 | 49 | 33 |
| 1986 | 1404 | 267 | 173 | 39 | 32 | 12 | 22 |
| 1987 | 6542 | 555 | 73 | 66 | 16 | 16 | 10 |
| 1988 | 1006 | 1194 | 150 | 29 | 28 | 9 | 9 |
| 1989 | 2015 | 148 | 366 | 47 | 10 | 13 | 5 |
| 1990 | 1320 | 504 | 54 | 125 | 22 | 7 | 8 |
| 1991 | 1604 | 364 | 279 | 25 | 51 | 12 | 4 |
| 1992 | 2008 | 513 | 122 | 110 | 11 | 24 | 8 |
| 1993 | 876 | 730 | 184 | 37 | 37 | 6 | 11 |
| 1994 | 1858 | 312 | 344 | 63 | 12 | 16 | 7 |
| 1995 | 1439 | 749 | 143 | 128 | 23 | 6 | 9 |
| 1996 | 884 | 566 | 287 | 52 | 46 | 10 | 6 |
| 1997 | 2004 | 254 | 229 | 87 | 18 | 19 | 6 |
| 1998 | 935 | 731 | 91 | 59 | 23 | 7 | 8 |
| 1999 | 598 | 309 | 267 | 21 | 15 | 7 | 4 |
| 2000 | 1633 | 196 | 87 | 73 | 6 | 6 | 4 |
| 2001 | 463 | 626 | 56 | 20 | 18 | 2 | 3 |
| 2002 | 985 | 160 | 189 | 15 | 6 | 6 | 1 |
| 2003 | 388 | 342 | 49 | 56 | 4 | 2 | 3 |
| 2004 | 526 | 118 | 97 | 15 | 15 | 1 | 2 |
| 2005 | 457 | 177 | 34 | 26 | 4 | 6 | 1 |
| 2006 | 705 | 180 | 59 | 12 | 11 | 2 | 3 |
| 2007 | 393 | 276 | 41 | 11 | 2 | 3 | 1 |
| 2008 | 246 | 154 | 89 | 8 | 3 | 1 | 1 |
| 2009 | 484 | 100 | 50 | 26 | 2 | 2 | 1 |
| 2010 | 388 | 207 | 33 | 13 | 8 | 1 | 1 |
| 2011 | 185 | 176 | 77 | 8 | 3 | 4 | 1 |
| 2012 | 389 | 72 | 74 | 16 | 1 | 2 | 2 |
| 2013 | 609 | 164 | 30 | 22 | 5 | 1 | 1 |
| 2014 | 413 | 223 | 64 | 10 | 8 | 3 | 1 |
| 2015 | 384 | 187 | 96 | 19 | 4 | 5 | 2 |
| 2016 | 210 | 183 | 87 | 37 | 8 | 4 | 4 |
| 2017 | 252 | 76 | 95 | 34 | 17 | 5 | 4 |
| 2018 | 187 | 110 | 32 | 40 | 13 | 9 | 3 |
| 2019* | 1452 | 78 | 58 | 15 | 23 | 9 | 7 |

*2019 values are standard errors on TSA-derived projections of population numbers.

Table 5.14. Cod in Division 6.a. TSA estimates for fishing mortality-at-age.

|  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 5.15. Cod in Division 6.a. Standard errors of TSA estimates for log fishing mortality-at-age.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.028 | 0.039 | 0.043 | 0.050 | 0.075 | 0.095 | 0.110 |
| 1982 | 0.024 | 0.029 | 0.042 | 0.052 | 0.074 | 0.120 | 0.151 |
| 1983 | 0.062 | 0.033 | 0.044 | 0.058 | 0.079 | 0.130 | 0.174 |
| 1984 | 0.036 | 0.041 | 0.049 | 0.063 | 0.090 | 0.140 | 0.160 |
| 1985 | 0.075 | 0.038 | 0.054 | 0.070 | 0.088 | 0.148 | 0.177 |
| 1986 | 0.042 | 0.040 | 0.044 | 0.057 | 0.082 | 0.152 | 0.141 |
| 1987 | 0.098 | 0.047 | 0.049 | 0.063 | 0.084 | 0.133 | 0.168 |
| 1988 | 0.107 | 0.047 | 0.045 | 0.054 | 0.083 | 0.119 | 0.139 |
| 1989 | 0.075 | 0.044 | 0.063 | 0.061 | 0.101 | 0.152 | 0.206 |
| 1990 | 0.090 | 0.052 | 0.054 | 0.075 | 0.080 | 0.122 | 0.129 |
| 1991 | 0.087 | 0.068 | 0.064 | 0.066 | 0.098 | 0.146 | 0.182 |
| 1992 | 0.063 | 0.056 | 0.083 | 0.083 | 0.096 | 0.116 | 0.160 |
| 1993 | 0.064 | 0.051 | 0.078 | 0.082 | 0.108 | 0.164 | 0.151 |
| 1994 | 0.070 | 0.049 | 0.064 | 0.089 | 0.114 | 0.154 | 0.182 |
| 1995 | 0.065 | 0.064 | 0.069 | 0.078 | 0.100 | 0.156 | 0.150 |
| 1996 | 0.079 | 0.067 | 0.088 | 0.086 | 0.104 | 0.183 | 0.201 |
| 1997 | 0.066 | 0.076 | 0.091 | 0.095 | 0.111 | 0.160 | 0.192 |
| 1998 | 0.091 | 0.070 | 0.098 | 0.096 | 0.091 | 0.163 | 0.168 |
| 1999 | 0.090 | 0.085 | 0.087 | 0.103 | 0.113 | 0.137 | 0.192 |
| 2000 | 0.092 | 0.077 | 0.098 | 0.096 | 0.122 | 0.162 | 0.210 |
| 2001 | 0.091 | 0.072 | 0.088 | 0.104 | 0.115 | 0.163 | 0.156 |
| 2002 | 0.112 | 0.084 | 0.086 | 0.100 | 0.147 | 0.184 | 0.278 |
| 2003 | 0.115 | 0.090 | 0.105 | 0.096 | 0.113 | 0.203 | 0.205 |
| 2004 | 0.102 | 0.088 | 0.096 | 0.115 | 0.142 | 0.231 | 0.239 |
| 2005 | 0.111 | 0.099 | 0.165 | 0.188 | 0.256 | 0.236 | 0.243 |
| 2006 | 0.096 | 0.143 | 0.167 | 0.220 | 0.208 | 0.246 | 0.244 |
| 2007 | 0.107 | 0.105 | 0.160 | 0.117 | 0.167 | 0.226 | 0.243 |
| 2008 | 0.086 | 0.101 | 0.125 | 0.143 | 0.141 | 0.222 | 0.231 |
| 2009 | 0.082 | 0.099 | 0.143 | 0.118 | 0.156 | 0.226 | 0.222 |
| 2010 | 0.065 | 0.085 | 0.135 | 0.133 | 0.120 | 0.192 | 0.203 |
| 2011 | 0.082 | 0.075 | 0.135 | 0.150 | 0.128 | 0.197 | 0.222 |
| 2012 | 0.079 | 0.067 | 0.098 | 0.113 | 0.121 | 0.176 | 0.190 |
| 2013 | 0.107 | 0.077 | 0.072 | 0.081 | 0.120 | 0.166 | 0.177 |
| 2014 | 0.064 | 0.063 | 0.106 | 0.063 | 0.084 | 0.126 | 0.150 |
| 2015 | 0.053 | 0.068 | 0.092 | 0.095 | 0.129 | 0.156 | 0.156 |
| 2016 | 0.078 | 0.060 | 0.099 | 0.086 | 0.087 | 0.151 | 0.162 |
| 2017 | 0.060 | 0.090 | 0.100 | 0.094 | 0.085 | 0.146 | 0.161 |
| 2018 | 0.066 | 0.081 | 0.132 | 0.131 | 0.137 | 0.184 | 0.197 |

Table 5.16. Cod in Division 6.a. TSA summary table.

| Year | TotalCatch (tonnes) Obs. | Pred. | SE | Landings (tonnes) Obs. | Pred. | SE | Discards (tonnes) Obs. | Pred | SE | Mean F(2- <br> 5) <br> Estimate | SE | SSB <br> (tonnes) <br> Estimate | SE | Recruitment (000s at age <br> 1) <br> Estimate | SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 198 | 24168 | 2421 | 954 | 23865 | 2405 | 956 | 303 | 167 | 95 | 0.672 | 0.02 | 41750 | 109 | 11028 | 126 |
| 198 | 22082 | 2093 | 802 | 21511 | 2014 | 810 | 571 | 793 | 18 | 0.666 | 0.02 | 39023 | 114 | 26437 | 113 |
| 198 | 21503 | 2117 | 698 | 21305 | 2084 | 687 | 197 | 329 | 15 | 0.761 | 0.03 | 34821 | 100 | 15317 | 119 |
| 198 | 21601 | 2147 | 794 | 21272 | 2086 | 807 | 329 | 613 | 24 | 0.865 | 0.03 | 32299 | 108 | 27103 | 123 |
| 198 | 19570 | 1841 | 706 | 18607 | 1780 | 713 | 963 | 613 | 16 | 0.948 | 0.03 | 25558 | 895 | 13406 | 127 |
| 198 | 12083 | 1275 | 503 | 11820 | 1217 | 475 | 263 | 575 | 20 | 0.783 | 0.03 | 20564 | 629 | 23663 | 140 |
| 198 | 21358 | 1932 | 115 | 18971 | 1755 | 921 | 2388 | 1774 | 77 | 0.886 | 0.03 | 21786 | 568 | 57984 | 654 |
| 198 | 20781 | 2063 | 100 | 20413 | 2029 | 989 | 368 | 338 | 18 | 0.864 | 0.03 | 27998 | 865 | 6830 | 100 |
| 198 | 19246 | 1825 | 863 | 17169 | 1685 | 764 | 2076 | 1406 | 50 | 0.951 | 0.04 | 24102 | 861 | 24723 | 201 |
| 199 | 12746 | 1247 | 500 | 12175 | 1224 | 491 | 571 | 228 | 90 | 0.785 | 0.04 | 19431 | 902 | 8515 | 132 |
| 199 | 11549 | 1065 | 992 | 10927 | 1017 | 939 | 622 | 478 | 21 | 0.803 | 0.05 | 16143 | 120 | 13386 | 160 |
| 199 | 10865 | 9851 | 101 | 9086 | 9211 | 967 | 1779 | 640 | 26 | 0.848 | 0.06 | 14020 | 120 | 22922 | 200 |
| 199 | 10453 | 1131 | 106 | 10314 | 1093 | 103 | 139 | 385 | 13 | 0.835 | 0.06 | 17042 | 131 | 9087 | 876 |
| 199 | 9588 | 1111 | 113 | 8928 | 1043 | 107 | 661 | 677 | 24 | 0.794 | 0.06 | 17693 | 144 | 18653 | 185 |
| 199 | 9580 | 1224 | 126 | 9439 | 1186 | 123 | 141 | 380 | 14 | 0.851 | 0.06 | 18040 | 150 | 14871 | 143 |
| 199 | 9489 | 1215 | 134 | 9427 | 1191 | 131 | 63 | 239 | 89 | 0.962 | 0.06 | 17111 | 152 | 6364 | 884 |
| 199 | 7533 | 1032 | 119 | 7034 | 9457 | 110 | 499 | 872 | 35 | 1.047 | 0.07 | 12106 | 121 | 23205 | 200 |
| 199 | 6252 | 9253 | 913 | 5714 | 8928 | 889 | 538 | 324 | 12 | 1.002 | 0.07 | 10273 | 907 | 6608 | 935 |
| 199 | 4270 | 6911 | 795 | 4201 | 6686 | 769 | 69 | 225 | 90 | 1.069 | 0.07 | 9078 | 860 | 4501 | 598 |
| 200 | 3798 | 6056 | 666 | 2977 | 4978 | 589 | 821 | 1078 | 31 | 1.034 | 0.07 | 6338 | 647 | 16469 | 163 |
| 200 | 2439 | 5265 | 606 | 2347 | 5055 | 586 | 92 | 211 | 72 | 1.091 | 0.07 | 6441 | 607 | 3566 | 463 |
| 200 | 2722 | 5103 | 577 | 2243 | 4627 | 526 | 480 | 476 | 16 | 1.111 | 0.08 | 5995 | 588 | 7436 | 985 |


| Year | TotalCatch (tonnes) Obs. | Pred. | SE | Landings (tonnes) Obs. | Pred. | SE | Discards (tonnes) Obs. | Pred | SE | Mean F(2- <br> 5) <br> Estimate | SE | SSB <br> (tonnes) <br> Estimate | SE | Recruitment (000s at age <br> 1) <br> Estimate | SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 200 | 1275 | 3661 | 487 | 1241 | 3519 | 466 | 34 | 142 | 59 | 1.078 | 0.07 | 4510 | 507 | 1978 | 388 |
| 200 | 612 | 2204 | 330 | 540 | 2017 | 301 | 72 | 187 | 71 | 0.992 | 0.07 | 2945 | 370 | 2947 | 526 |
| 200 | 552 | 1828 | 283 | 511 | 1707 | 264 | 41 | 121 | 50 | 1.150 | 0.11 | 2309 | 281 | 1660 | 457 |
| 200 | 954 | 1579 | 271 | 488 | 493 | 114 | 465 | 1086 | 23 | 1.073 | 0.11 | 1790 | 219 | 5291 | 705 |
| 200 | 2474 | 1815 | 220 | 595 | 487 | 64 | 1880 | 1328 | 20 | 1.037 | 0.07 | 2299 | 185 | 2088 | 393 |
| 200 | 1377 | 1587 | 192 | 682 | 583 | 90 | 695 | 1004 | 18 | 0.913 | 0.07 | 2488 | 210 | 1424 | 246 |
| 200 | 1353 | 1506 | 161 | 408 | 475 | 48 | 945 | 1031 | 16 | 0.903 | 0.07 | 2124 | 169 | 3681 | 484 |
| 201 | 1344 | 1612 | 194 | 559 | 555 | 53 | 785 | 1057 | 18 | 0.858 | 0.06 | 2385 | 182 | 3642 | 388 |
| 201 | 2124 | 1883 | 203 | 454 | 474 | 60 | 1670 | 1409 | 19 | 1.013 | 0.06 | 2769 | 197 | 1586 | 185 |
| 201 | 1632 | 1528 | 155 | 466 | 488 | 49 | 1166 | 1041 | 15 | 0.734 | 0.05 | 2715 | 177 | 2407 | 389 |
| 201 | 1501 | 1268 | 118 | 299 | 392 | 36 | 1202 | 876 | 12 | 0.624 | 0.05 | 2347 | 134 | 4986 | 609 |
| 201 | 1668 | 1480 | 181 | 357 | 416 | 43 | 1311 | 1064 | 17 | 0.532 | 0.04 | 3042 | 205 | 3592 | 413 |
| 201 | 1752 | 2091 | 234 | 770 | 718 | 89 | 983 | 1372 | 20 | 0.631 | 0.05 | 4279 | 268 | 3337 | 384 |
| 201 | 1745 | 2037 | 216 | 892 | 805 | 78 | 852 | 1232 | 20 | 0.548 | 0.04 | 4642 | 281 | 1282 | 210 |
| 201 | 2363 | 1969 | 192 | 795 | 873 | 70 | 1569 | 1096 | 18 | 0.600 | 0.05 | 4196 | 274 | 1579 | 252 |
| 201 | 1890 | 1791 | 148 | 1129 | 1128 | 75 | 760 | 663 | 12 | 0.700 | 0.08 | 3478 | 265 | 826 | 187 |
| 201 | NA | 1402 | 198 | NA | 743 | 124 | NA | 659 | 14 | 0.668 | 0.11 | 2357 | 313 | 6404 | 145 |
| Min | 552 | 1268 | 118 | 299 | 392 | 36 | 34 | 121 | 50 | 0.532 | 0.02 | 1790 | 134 | 826 | 185 |
| GM | 4562 | 5501 | 474 | 3028 | 3681 | 317 | 466 | 576 | 16 | 0.852 | 0.05 | 8137 | 528 | 6374 | 761 |
| AM | 8113 | 8677 | 609 | 7367 | 7953 | 541 | 746 | 724 | 19 | 0.869 | 0.06 | 12788 | 684 | 10642 | 105 |
| Max | 24168 | 2421 | 134 | 23865 | 2405 | 131 | 2388 | 1774 | 77 | 1.150 | 0.11 | 41750 | 152 | 57984 | 654 |

*Estimates for 2019 are TSA projections.

Table 5.17. Cod in Division 6.a.. Input values for short-term forecast. Note that LSel and LWt refer to the landings and DSel and DCWt refer to the discards. Numbers in thousands; Weights in kg.

2019

| Age | N | M | Mat | PF | PM | SWt | LSel | LWt | DSel | DWt |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2349 | 0.541 | 0 | 0 | 0 | 0.240 | 0.103 | 0.810 | 0.130 | 0.234 |
| 2 | 389 | 0.387 | 0.52 | 0 | 0 | 1.087 | 0.456 | 1.087 | 0.018 | 1.042 |
| 3 | 344 | 0.307 | 0.86 | 0 | 0 | 2.298 | 0.698 | 2.298 | 0.00028 | 2.208 |
| 4 | 88 | 0.263 | 1 | 0 | 0 | 3.886 | 0.767 | 3.886 | 0 | 3.634 |
| 5 | 120 | 0.238 | 1 | 0 | 0 | 5.281 | 0.861 | 5.281 | 0 | 4.767 |
| 6 | 40 | 0.224 | 1 | 0 | 0 | 6.130 | 0.944 | 6.130 | 0 | 6.841 |
| 7 | 27 | 0.212 | 1 | 0 | 0 | 8.242 | 0.905 | 8.242 | 0 | 2.118 |

2020

| Age | N | M | Mat | PF | PM | SWt | LSel | LWt | DSel | DWt |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2349 | 0.541 | 0 | 0 | 0 | 0.240 | 0.103 | 0.810 | 0.130 | 0.234 |
| 2 | 1083 | 0.387 | 0.52 | 0 | 0 | 1.087 | 0.456 | 1.087 | 0.018 | 1.042 |
| 3 | 165 | 0.307 | 0.86 | 0 | 0 | 2.298 | 0.698 | 2.298 | 0.00028 | 2.208 |
| 4 | 126 | 0.263 | 1 | 0 | 0 | 3.886 | 0.767 | 3.886 | 0 | 3.634 |
| 5 | 32 | 0.238 | 1 | 0 | 0 | 5.281 | 0.861 | 5.281 | 0 | 4.767 |
| 6 | 40 | 0.224 | 1 | 0 | 0 | 6.130 | 0.944 | 6.130 | 0 | 6.841 |
| 7 | 21 | 0.212 | 1 | 0 | 0 | 8.242 | 0.905 | 8.242 | 0 | 2.118 |

2021

| Age | N | M | Mat | PF | PM | SWt | LSel | LWt | DSel | DWt |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2349 | 0.541 | 0 | 0 | 0 | 0.240 | 0.103 | 0.810 | 0.130 | 0.234 |
| 2 | 1083 | 0.387 | 0.52 | 0 | 0 | 1.087 | 0.456 | 1.087 | 0.018 | 1.042 |
| 3 | 458 | 0.307 | 0.86 | 0 | 0 | 2.298 | 0.698 | 2.298 | 0.00028 | 2.208 |
| 4 | 60 | 0.263 | 1 | 0 | 0 | 3.886 | 0.767 | 3.886 | 0 | 3.634 |
| 5 | 45 | 0.238 | 1 | 0 | 0 | 5.281 | 0.861 | 5.281 | 0 | 4.767 |
| 6 | 11 | 0.224 | 1 | 0 | 0 | 6.130 | 0.944 | 6.130 | 0 | 6.841 |
| 7 | 19 | 0.212 | 1 | 0 | 0 | 8.242 | 0.905 | 8.242 | 0 | 2.118 |

Table 5.18. Cod in Division 6.a. Single-option output of the short-term forecast ( $F=$ mean $F 2016$-2018, rescaled to 2018). Numbers in thousands, weights in tonnes.

2019

| Age | F | LandNos | Yield | DF | DiscNos | DYield | StockNos | Biomass | SSNos | SSB |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.233 | 168 | 136 | 0.130 | 213 | 50 | 2349 | 565 | 0 | 0 |
| 2 | 0.475 | 119 | 130 | 0.018 | 5 | 5 | 389 | 423 | 202 | 220 |
| 3 | 0.699 | 151 | 348 | 0.00028 | 0 | 0 | 344 | 790 | 296 | 679 |
| 4 | 0.767 | 42 | 165 | 0 | 0 | 0 | 88 | 344 | 88 | 344 |
| 5 | 0.861 | 63 | 331 | 0 | 0 | 0 | 120 | 634 | 120 | 634 |
| 7 | 0.944 | 22 | 137 | 0 | 0 | 0 | 40 | 246 | 40 | 246 |
| 7 | 0.905 | 15 | 120 | 0 | 0 | 0 | 27 | 221 | 27 | 221 |
| Total | 0.700 | 580 | 1366 | 0.0046 | 218 | 55 | 3357 | 3223 | 773 | 2344 |

2020

| Age | F | LandNos | Yield | DF | DiscNos | DYield | StockNos | Biomass | SSNos | SSB |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.233 | 168 | 136 | 0.130 | 213 | 50 | 2349 | 565 | 0 | 0 |
| 2 | 0.475 | 331 | 360 | 0.018 | 13 | 14 | 1083 | 1177 | 563 | 612 |
| 3 | 0.699 | 72 | 167 | 0.00028 | 0 | 0 | 165 | 378 | 142 | 325 |
| 4 | 0.767 | 60 | 234 | 0 | 0 | 0 | 126 | 489 | 126 | 489 |
| 5 | 0.861 | 16 | 87 | 0 | 0 | 0 | 32 | 167 | 32 | 167 |
| 6 | 0.944 | 22 | 137 | 0 | 0 | 0 | 40 | 245 | 40 | 245 |
| 7 | 0.905 | 12 | 95 | 0 | 0 | 0 | 21 | 175 | 21 | 175 |
| Total | 0.700 | 681 | 1216 | 0.0046 | 226 | 63 | 3816 | 3196 | 924 | 2013 |

2021

| Age | F | LandNos | Yield | DF | DiscNos | DYield | StockNos | Biomass | SSNos | SSB |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.233 | 168 | 136 | 0.130 | 213 | 50 | 2349 | 565 | 0 | 0 |
| 2 | 0.475 | 331 | 360 | 0.018 | 13 | 14 | 1083 | 1177 | 563 | 612 |
| 3 | 0.699 | 202 | 463 | 0.00028 | 0 | 0 | 458 | 1052 | 394 | 904 |
| 4 | 0.767 | 29 | 112 | 0 | 0 | 0 | 60 | 234 | 60 | 234 |
| 5 | 0.861 | 23 | 124 | 0 | 0 | 0 | 45 | 237 | 45 | 237 |
| 6 | 0.944 | 6 | 36 | 0 | 0 | 0 | 11 | 65 | 11 | 65 |
| 7 | 0.905 | 11 | 87 | 0 | 0 | 0 | 19 | 160 | 19 | 160 |
| Total | 0.700 | 770 | 1318 | 0.0046 | 226 | 64 | 4025 | 3490 | 1092 | 2212 |

Table 5.19. Cod in Division 6.a. Management options table (plus table covering the interval between lower and upper bounds of $\mathrm{F}_{\mathrm{ms}}$ ). Weights in tonnes.

| Basis | $\begin{aligned} & \text { Catc } \\ & \text { h } \end{aligned}$ | Wanted.catc h | Unwanted.catch | Ftotal | F.wante d | F.unwanted | SSB | Perc. SSB | Perc.TA <br> C |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fx 0 | 0 | 0 | 0 | 0 | NA | NA | 3765 | 87 | -100 |
| F $\times 1$ | 1279 | 1216 | 63 | 0.70010 | 0.69552 | 0.00458 | 2212 | 9.9 | -30 |
| $\mathrm{F}_{\text {MSY }}$ | 623 | 595 | 28 | 0.29000 | 0.28810 | 0.00190 | 3002 | 49 | -66 |
| $\mathrm{F}_{\mathrm{pa}}$ | 1064 | 1013 | 51 | 0.55000 | 0.54640 | 0.00360 | 2469 | 23 | -42 |
| $\mathrm{F}_{\text {lim }}$ | 1371 | 1302 | 69 | 0.77000 | 0.76496 | 0.00504 | 2103 | 4.5 | -25 |
| $F_{\text {msy.up }}$ | 839 | 800 | 39 | 0.41000 | 0.40732 | 0.00268 | 2741 | 36 | -54 |
| $\mathrm{F}_{\text {MSY.Iow }}$ | 446 | 426 | 20 | 0.20000 | 0.19869 | 0.00131 | 3218 | 60 | -75 |
| scaled. $\mathrm{F}_{\text {MSY }}$ | 70 | 67 | 3 | 0.02919 | 0.02900 | 0.00019 | 3679 | 83 | -96 |
| $\begin{aligned} & \text { scaled. } \begin{array}{l} \text { MSY.up- } \\ \text { per } \end{array} \end{aligned}$ | 98 | 94 | 4 | 0.04127 | 0.04100 | 0.00027 | 3644 | 81 | -95 |
| scaled. FMsy.lowe | 48 | 46 | 2 | 0.02013 | 0.02000 | 0.00013 | 3705 | 84 | -97 |



Figure 5.1. Distribution of Scottish reported landings by statistical rectangle by year.


Figure 5.2. Cod in Division 6.a. Estimates of underreporting and area misreporting of cod caught in ICES Division 6.a by Scottish vessels. Negative values of area misreporting indicate a net balance of misreporting into Division 6.a from other areas.


Figure 5.3. Cod in Division 6.a. Amounts landed by métier (kg) in 2018 as submitted to InterCatch.


Figure 5.4. Cod in Division 6.a. Amounts discarded by métier (kg) in 2018 as submitted to InterCatch.


Figure 5.5. Cod in Division 6.a. Discard rates before allocations within InterCatch.
cod.27.6a CatchWt


Figure 5.6. Cod in Division 6.a. Landings (grey), imported (black) and raised (red) discards of all fleets after allocations within InterCatch.


Figure 5.7. Cod in Division 6.a. Number-at-age constituted by sampled and unsampled landings and sampled and raised (unsampled) discards after allocations within InterCatch. Minor amounts ( $\mathbf{3 5} \mathbf{~ k g}$ ) of BMS landings were also recorded.


Figure 5.8. Cod in Division 6.a. Landings and discards estimates by weight, as used by the WG. Values are totals for fish aged 1 to 7+.


Figure 5.9. Cod in Division 6.a. Discard proportion (of total catch) by weight. Includes fish aged 1 to $7+$.


Figure 5.10. Cod in Division 6.a. Discard proportion by number.


Figure 5.11. Cod in Division 6.a. Catch-at-age in numbers by year. Pink: discards, blue: landings.


Figure 5.12. Cod in Division 6.a. Mean weights-at-age in landings and discards.


Figure 5.13. Cod in Division 6.a. Catch numbers for fish aged at 1+ per haul resulting from quarter four Irish ground fish survey (IRGFS-WIBTS-Q4). Values are standardised to 60 minutes towing. Zero shown as a black + symbol.


Figure 5.14. Cod in Division 6.a. Cpue numbers for fish aged at 1+ per tow resulting from Scottish quarter one survey (UK-SCOWCGFS-Q1) in red and (UK-SCOWCGFS-Q4) in blue. Numbers are standardised to 30 minutes towing. Green polygons are areas closed to fishing.


Figure 5.15. Cod in Division 6.a. Natural mortality-at-age based on mean weight-at-age and mortality-weight relationship. Solid horizontal lines show the time averaged values at each age used in the assessment. Dotted horizontal line shows value of 0.2 previously used at all ages in all years.


Figure 5.16. Cod in Division 6.a. Catch curves from commercial catch-at-age data.



Figure 5.17. Cod in Division 6.a. Log catch (landings + discards) curve gradient plot using WG commercial catch-at-age data over different age ranges.


Figure 5.18. Cod in Division 6.a. Mean standardised catch-at-age proportions by number.


Figure 5.19. Cod in Division 6.a. Log mean standardised index values, by year, (left) and cohort (right) from Scottish quarter one ground fish survey (ScoGFS-WIBTS-Q1); ages 1-6. Survey finished in 2010.


Figure 5.20. Cod in Division 6.a. Log catch curves from Scottish quarter one ground fish survey (ScoGFS-WIBTS-Q1); ages 1-6. Survey finished in 2010.

## ScoGFS-WIBTS-Q1



Figure 5.21. Cod in Division 6.a. Within-survey correlations for the Scottish quarter one ground fish survey (ScoGFS-WIBTS-Q1), comparing index values at different ages for the same cohorts. The straight line in a linear regression. Survey finished in 2010.


Figure 5.22. Cod in Division 6.a. Log catch curve gradient plot using ScoGFS-WIBTS-Q1 index data. Solid line shows time-series of gradient of linear fit to curve over the age range $2-5$, dashed line over the ages 2-4 and dotted line over the ages 3-5. Last cohort shown was at-age 5 in 2010, the last year of the ScoGFS-WIBTS-Q1 survey.


Figure 5.23. Cod in Division6a. Log mean standardised index values by year (left) and cohort (right) from ScoGFS-WIBTS-Q4.

log index

Figure 5.24. Cod in Division 6.a. Within survey correlations for ScoGFS-WIBTS-Q4 survey, comparing index values at different ages for the same cohorts. The solid line is a linear regression. Insufficient age 6 fish are caught to enable scatterplots to be constructed.


Figure 5.25. Cod in Division 6.a. Log mean standardised index values, by year (left) and cohort (right) from Irish quarter four ground fish survey (IRGFS-WIBTS-Q4); ages 1-3. Survey started in 2003.


Figure 5.26. Cod in Division 6.a. Log catch curves from Irish quarter four ground fish survey (IRGFS-WIBTSQ4); ages 1-4. Survey started in 2003.

IGFS-WIBTS-Q4


Figure 5.27. Cod in Division 6.a. Within-survey correlations for the Irish quarter four ground fish survey (IRGFS-WIBTS-Q4), comparing index values at different ages for the same cohorts. The straight line is a linear regression.


Figure 5.28. Cod in Division 6.a. Log mean standardised index values -by year (left) and cohort (right); from Scottish quarter one ground fish survey UK-SCOWCGFS-Q1; ages 1-6.


Figure 5.29. Cod in Division 6.a. Log catch curves from new Scottish quarter one ground fish survey (UK-SCOWCGFS-Q1); ages 1-7. Survey started in 2011.

UKSGFS-WIBTS-Q1

log index

Figure 5.30. Cod in Division 6.a. Within survey scatterplots from new Scottish quarter one ground fish survey (UK-SCOWCGFS-Q1), comparing index values at different ages for the same cohorts. The straight line in a linear regression.


Figure 5.31. Cod in Division 6.a. Log mean standardised index values by year (left) and cohort (right) from Scottish quarter four ground fish survey UK-SCOWCGFS-Q4); ages 0-6.


Figure 5.32. Cod in Division 6.a. Log catch curves from new Scottish quarter four ground fish survey (UK-SCOWCGFS-Q4).


Figure 5.33. Cod in Division 6.a. Within survey scatterplots from new Scottish quarter four ground fish survey (UK-SCOWCGFS-Q4), comparing index values at different ages for the same cohorts. The straight line in a linear regression.


Figure 5.34. Cod in Division 6a. Comparison of survey indices by age. Irish Q4 survey (IRGFS.Q4) is compared to the current Scottish surveys. Values are mean standardised over the time period in common (20112018).


Figure 5.35. Cod in Division 6.a. TSA final run. Standardised residuals at-age for landings (update assessment).



Figure 5.36. Cod in Division 6.a. TSA final run. Residuals at-age plots for landings (upper) and discards (lower). Final assessment.





Figure 5.37. Cod in Division 6.a. TSA final run. Residuals at-age plots for ScoGFS-WIBTS-Q1, UK-SCOWCGFS-Q1, ScoGFS-WIBTS-Q4, UK-SCOWCGFS-Q4 \& IRGFS-WIBTS-Q4 (upper to lower plots).


Figure 5.38. Cod in Division 6.a. Observed (points) and fitted (red lines with 95\% CI indicated by grey bands) for the proportion discarded by age. Commercial data from 2006 (indicated by an unfilled circle) are not included in the assessment. Note that the plot also shows the TSA projection of discards for 2019.


Figure 5.39. Cod in Division 6.a. Summary plot of final TSA run. Stock summary from final TSA assessment. Red lines (or points) give best estimates, grey bands (or lines) give approximate pointwise $95 \%$ confidence intervals, and black points give observed values. Commercial data from 2006 (indicated by an unfilled circle) are not included in the assessment. Note that the plot also shows the TSA projection for 2019.


Figure 5.40. Cod in Division 6.a. Retrospective plots of final TSA run.


Figure 5.41. Cod in Division 6.a. TSA final run. Stock-recruit relationship. Numbers indicate year class.
a)

b)


Figure 5.42. Cod in Division 6.a. a) Comparison of estimated mean F and STECF effort data; b) Partial mean F attributed to landings and discards. Horizontal lines represent $\mathrm{F}_{\text {lim }}$ (solid), $\mathrm{F}_{\mathrm{pa}}$ (dashed) and $\mathrm{Fmsy}^{\text {(dotted) values }}$ for the stock.


Figure 5.43. Cod in Division 6.a. Percentage contribution to landings yield in 2020 and SSB in 2021 by recruitment year (not year class).


Figure 5.44. Cod in Division 6.a. Comparison of recent assessments (WGCSE 2018, IBP 2019 \& WGCSE 2019).

## 4 Cod (Gadus morhua) in Division 6.b (Rockall)

## Assessment in 2017

In 2017, the update assessment and advice followed the agreed procedures for category 6.2.0 of ICES RGLIFE data-limited stock (DLS) advice rules (ICES, 2017a) as set out in the stock annex. For stocks without information on abundance or exploitation, ICES considers that a precautionary reduction of catches should be implemented.

Given there are conflicting signals from the Irish otter-trawl and Scottish TR1 fleet effort and lpue series and that survey catch rates at Rockall remain too low to provide quantitative information on abundance the ICES advice is to apply the precautionary buffer (last applied in 2012).

ICES advice applicable in 2018-2020
ICES advises that when the precautionary approach is applied, catches should be no more than 14 tonnes in each of the years 2018,2019, and 2020. ICES cannot quantify the corresponding landings.

ICES advice applicable in 2016-2017
ICES advises that when the precautionary approach is applied, landings should be no more than 17 tonnes in each of the years 2016 and 2017. ICES cannot quantify the corresponding total catches.

## ICES approach to data-limited stocks

For data-limited stocks without information on abundance or exploitation ICES considers that a precautionary reduction of catches should be implemented, unless there is ancillary information clearly indicating that the current level of exploitation is appropriate for the stock.

### 4.1 General

## Management applicable to 2018-2020

The TAC for cod at Rockall covers ICES Division 6.b, EU and international waters of Division $5 . \mathrm{b}$ west of $12^{\circ} 00^{\prime} \mathrm{W}$ and subareas 12 and 14 . The following is applicable to 2018-2020:

| Species:Cod <br> Gadus morhua | Zone:VIb; Union and international waters of Vb west of <br> $12^{\circ} 00^{\prime} W$ and of XII and XIV <br> (COD/5W6-14) |  |
| :--- | :--- | :--- | :--- |
| Belgium | 0 |  |
| Germany | 1 |  |
| France | 12 |  |
| Ireland | 16 |  |
| United Kingdom | 45 |  |
| Union | 74 | Precautionary TAC |
| TAC | 74 |  |

## The fishery in 2018

No specific information is available for 2018. Cod at Rockall are taken as a bycatch in fisheries for other species such as haddock and anglerfish.

### 4.2 Data

Official landings data for cod in 6.b are shown by nation in Table 4.2.1 and Figure 4.2.1. Total reported landings were 64.1 tonnes in 2018. There were minor updates to landings from previous years as well as slight modifications to the Irish effort time series due to cleaning of historic logbook data (Table 6.2). In the past, official landings have shown very high interannual variation and it is not known whether these are a true reflection of removals.

Landings data have been uploaded to InterCatch for 2018. In addition, some landings age compositions and discard data were also uploaded to InterCatch. Data uploaded to InterCatch are shown below.

| Country | Discards (t) | Landings (t) |
| :--- | :---: | :---: |
| Ireland |  | 14.2 |
| Norway | 71.9 | 0.6 |
| UK (Scotland) | 49.2 |  |
| Grand Total | 64.1 |  |

In recent years only limited discard data have been submitted to InterCatch for this stock. Discarded weight has been submitted for the Scottish demersal otter trawl fleet for the years 2014-2018 however there is high interannual variability in the estimated discard rate for this fleet ( $0-60.90 \%$ ). Discard information has also been provided by Ireland in 2016 and 2017 for the demersal otter trawl fleet 100-119 mm mesh size which has more consistent discard rate estimates ( $2.25 \%$ and $3.68 \%$ ) although no discard information was submitted for this fleet in 2018. This means that it is difficult to determine an appropriate discard rate for use in the provision of catch advice.

Irish and Scottish landings, effort and lpue are presented in Figures 6.2 and 6.3 and Tables 6.2 and 6.4. Figure 6.2 shows a large decline in the Irish lpue between 1995 and 2003 followed by relatively stable values at a level much lower than at the start of the time-series. In 2017 there was a large increase in effort for this fleet exceeding the previous time-series maximum. This has fallen in 2018 but remains above the 1995-2016 maximum. The recording of Scottish hour's fished data is not mandatory in the log
sheets and the data are incomplete. Scottish otter-trawl fleet data are therefore in units of $\mathrm{kg} / \mathrm{kWday}$. The Scottish time-series is much shorter and relatively noisier. Whilst there were marked increases in lpue in 2015 and 2016, given the magnitude of increase it seems unlikely to be completely attributable to an increase in stock size (an almost five-fold increase over two years). 2018 has seen a moderate increase in landings with no change in fishing effort which has led to an increase in lpue. The increase in Irish otter-trawl effort since 2010 has been anecdotally attributed to increases in the squid fishery in which cod is not a target or common bycatch species. This brings into question the usefulness of this lpue series as an indicator of cod abundance.

Survey catch rates of cod at Rockall remain low and are therefore unlikely to provide a reliable index of abundance (Table 6.4).

Catches of cod (both survey and commercial) are too low to support the collection of the necessary information for an assessment of stock status.

### 4.3 References

ICES. 2017a. Advice basis. In Report of the ICES Advisory Committee, 2017. ICES Advice 2017,
Book 1, Section 1.2.

Table 6.1. Cod in Division 6.b (Rockall). Official catch statistics.

| Country | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe <br> Islands | 18 | - | 1 | - | 31 | 5 | - | - | - | 1 | - | - | - | - | - | - | - |
| France | 9 | 17 | 5 | 7 | 2 | - | - | - | - | - | - | - | - | - | - | - | + |
| Germany | - | 3 | - | - | 3 | - | - | 126 | 2 | - | - | - | 10 | 22 | 3 | 11 | 1 |
| Ireland | - | - | - | - | - | - | 400 | 236 | 235 | 472 | 280 | 477 | 436 | 153 | 227 | 148 | 119 |
| Norway | 373 | 202 | 95 | 130 | 195 | 148 | 119 | 312 | 199 | 199 | 120 | 92 | 91 | 55 | 52 | 85 | 152 |
| Portugal | - | - | - | - | - | - | - | - | - | - | - | - | - | 5 | - | - | - |
| Russia | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 7 |
| Spain | 241 | 1200 | 1219 | 808 | 1345 | - | 64 | 70 | - | - | - | 2 | 5 | 1 | 6 | 4 | 3 |
| UK (E. \& W. \& N.I.) | 161 | 114 | 94 | 69 | 56 | 131 | 8 | 23 | 26 | 103 | 25 | 90 | 23 | 20 | 32 | 22 | 4 |
| UK (Scotland) | 221 | 437 | 187 | 284 | 254 | 265 | 758 | 829 | 714 | 322 | 236 | 370 | 210 | 706 | 341 | 389 | 286 |
| UK | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Total | 1023 | 1973 | 1601 | 1298 | 1886 | 549 | 1349 | 1596 | 1176 | 1097 | 661 | 1031 | 775 | 962 | 661 | 659 | 572 |

Table 6.1. Continued. Cod in Division 6.b (Rockall). Official catch statistics.

| Country | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe <br> Islands | - | - | - | - | - | - | - | - | 3 | 5 | - | - | - | - | - | + | - | - |
| France | - | + | + | - | - | - | - | - | - | - | - | + | - | - | - | - | - | - |
| Germany | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Ireland | 40 | 18 | 11 | 7 | 12 | 23 | 24 | 41 | 20 | 6 | 12 | 1 | 2 | 6 | 5 | 15 | 17 | 14 |
| Norway | 89 | 28 | 25 | 23 | 7 | 7 | 12 | 12 | 25 | 27 | 49 | 11 | 3 | 1 | 18 | 11 | 3 | 1 |
| Portugal | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Russia | 26 | - | - | - | - | - | - | - | 1 | - | - | - | - | - | - | - | - | - |
| Spain | 1 | + | 6 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| UK (E. \& W. \& N.I.) | 2 | 2 | 3 | - | - | + | - | - | - | - | - | - | - | - | - | - | - | - |
| UK (Scotland) | 176 | 67 | 57 | 45 | 43 | 28 | 26 | 41 | 48 | 23 | 37 | 11 | 9 | - | - | - | - | - |
| UK | - | - | - | - | - | - | - | - | - | - | - | - | - | 10 | 18 | 37 | 38 | 49 |
| Total | 334 | 115 | 102 | 75 | 62 | 58 | 62 | 94 | 97 | 61 | 98 | 23 | 14 | 17 | 41 | 62 | 58 | 64 |

* Preliminary.

Table 6.2. Cod in 6.b. Landings, effort and lpue data from Irish otter-trawl fleet.

| Year | Landings tonnes | Effort '000s Hrs | LPUE Kg/Hr |
| :---: | :---: | :---: | :---: |
| 1995 | 414.9 | 9.1 | 45.39 |
| 1996 | 402 | 7.2 | 55.68 |
| 1997 | 130.5 | 7.2 | 18.2 |
| 1998 | 207.1 | 7.3 | 28.23 |
| 1999 | 137.8 | 8.79 | 15.88 |
| 2000 | 101.1 | 9.9 | 10.23 |
| 2001 | $33.3$ | 7.2 | 4.6 |
| 2002 | 16.2 | 2.6 | 6.18 |
| 2003 | 9.9 | 4.5 | 2.18 |
| 2004 | 6.9 | 2.2 | 3.08 |
| $2005$ | 8.8 | 3.3 | 2.68 |
| 2006 | 22.2 | 5.9 | 3.76 |
| $2007$ | 24.2 | 6.6 | 3.68 |
| 2008 | 41.6 | 9.9 | 4.21 |
| 2009 | 21.7 | 4.4 | 4.97 |
| $2010$ | $7.5$ | 3.3 | 2.3 |
| 2011 | 10.2 | 2.5 | 4.01 |
| 2012 | 1 | 3.2 | 0.31 |
| $2013$ | $1.8$ | 3.8 | 0.46 |
| 2014 | 5.6 | 4.2 | 1.34 |
| 2015 | 5.1 | 4.7 | 1.07 |
| 2016 | 16.4 | 6.2 | 2.65 |
| 2017 | 17.3 | 14.9 | 1.16 |
| 2018 | 13.3 | 11.8 | 1.13 |

Table 6.3. Cod in 6.b. Landings, effort and lpue data from the Scottish TR1 fleet.

| year | Inds(t) | eff(kwdays) | LPUE(kg/kwday) |
| :--- | :--- | :--- | :--- |
| 2003 | 64.09 | 2504466 | 0.0256 |
| 2004 | 39.76 | 1842103 | 0.0216 |
| 2005 | 42.98 | 1217357 | 0.0353 |
| 2006 | 28.25 | 1011354 | 0.0279 |
| 2007 | 25.98 | 1060551 | 0.0245 |
| 2008 | 40.29 | 1124197 | 0.0358 |
| 2009 | 47.76 | 1631239 | 0.0293 |
| 2010 | 22.65 | 1744452 | 0.0130 |
| 2011 | 36.54 | 1565753 | 0.0233 |
| 2012 | 10.78 | 901552 | 0.0120 |
| 2013 | 9.09 | 532767 | 0.0171 |
| 2014 | 9.70 | 668665 | 0.0145 |
| 2015 | 19.92 | 563098 | 0.0354 |
| 2016 | 34.01 | 514486 | 0.0661 |
| 2017 | 49.25 | 794571 | 0.0475 |
| 2018 | 794017 | 0.062 |  |

Table 6.4. Cod in 6.b. Survey data made available to the WG: Scottish Q3 ground fish survey ((Rock-WIBTSQ3)). Catch rates are given as number per 10 hours.

| YEAR | Effort <br> (10 <br> Hours) | Age 0 | A GE 1 | AGE 2 | AGE 3 | $\begin{aligned} & \text { A GE } \\ & 4 \end{aligned}$ | AGE 5 | AGE 6 | $\begin{aligned} & \text { A GE } \\ & 7 \end{aligned}$ | AGE 8 | $\begin{aligned} & \text { A GE } \\ & 9 \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 10 | 0 | 0.493 | 0.493 | 0 | 0 | 0 | 0 | 0 | 0.403 | 0 |
| 2014 | 10 | 0 | 0.279 | 0.894 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 10 | 0 | 0 | 0.922 | 0.307 | 0 | 0 | 0 | 0 | 0 | 0.307 |
| 2016 | 10 | 0 | 0 | 0.269 | 0.538 | 0.538 | 0 | 0 | 0.269 | 0 | 0 |
| 2017 | 10 | 0 | 0 | 0 | 0 | 0.922 | 1.062 | 0 | 0 | 0 | 0 |
| 2018 | 10 | 0 | 0 | 0.307 | 0.614 | 0.307 | 0.307 | 0 | 0 | 0 | 0 |



Figure 6.1. Cod in Division 6.b. Total of official catch by nation. Values for 2018 are provisional.


Figure 6.2. Cod in Division 6.b. Landings, effort and lpue (kg/hr) from the Irish Otter-trawl fleet.


Figure 6.3. Cod in Division 6.b. Landings, effort and lpue (Kg/kWday) from the Scottish TR1 fleet.

## 5 Cod in 7.a (Irish Sea)

Situated between Ireland and Great Britain the Irish Sea (7.a) is connected by to the Celtic Sea (7.g) at its southern extreme by the St George's Channel and in north is linked to sea region West of Scotland (6.a) by the Northern Channel. The average depth is 50 m but the area is contrasted between a deeper channel, in the west, and shallower bays in the east. The channel has a maximum depth exceeding 275 m whilst the eastern bays have depths less than 50 m . Distinct habitat patches result from a combination of bathymetry, topographical features and hydrography. The sea bed of the eastern Irish Sea is dominated by fine sediment plains with some small areas of areas of mud habitat, the fine sediments graduate to more coarse material in central areas. A large well defined deep-water mud basin is located in the northwestern region in close to the Northern Irish and Irish coast.

Irish Sea fisheries are predominantly demersal trawling and seining with demersal trawling for Nephrops dominating effort with vessels using mesh in the range $70-99 \mathrm{~mm}$. Effort using fishing gear with $\geq 100 \mathrm{~mm}$ mesh sizes is currently at a low level compared to historic activity, a considerable decline in effort was observed between 2003 and 2007 and has continued. The species composition of catches by vessels in using $\geq 100 \mathrm{~mm}$ mesh consists of primarily haddock, with lower quantities of hake. At present there is no commercial towed gear fishery for cod permitted. Beam trawls areoperating within the Irish Sea with mesh sizes in the range $80-119 \mathrm{~mm}$, targeting sole, plaice, and rays. A seasonal pelagic and gillnet herring fishery operates in late summerearly autumn in the pre and post spawning period. Dredge fisheries target king and queen scallops, with king scallops in coastal areas with the queen scallop fishery operating in the central area south of the Isle of Man, to a lesser extent queen scallops are also targeted using trawl nets, during the late summer when swimming activity is most pronounced.

## Type of assessment

An ICES category 3 assessment based on a biomass trend was used as the full analytical assessment, benchmarked at ICES WKIrish3 (ICES, 2017a), did provide a retrospective bias believed to be too high.

## ICES advice applicable to 2018 and 2019

ICES advised on the basis of the MSY and precautionary approaches that there should be no directed fisheries, and bycatch and discards should be minimized in 2017. A TAC based on the MSY approach was advised for 2018 and 2019.

### 5.1 General

## Stock description and management units

The stock and the management unit are both ICES Division 7.a (Irish Sea).

TACs and quotas set for 2018

| Zone 7a (COD/07A) | Analytical TAC | Weight tonnes | Landed |
| :--- | :--- | :---: | :---: |
| Belgium | 9 | 1.9 |  |
| France | 25 | 0.05 |  |
|  | Theland | 459 | 84.6 |
|  | United Kingdom | 2 | 128.5 |
| EU | 200 | 214.9 |  |

Management of this cod fishery is by TAC, days-at-sea limits and technical measures. Technical regulations in force in the Irish Sea, including those associated with the cod recovery plan since 2000, are described in Section 7.2 and 7.10.

## Fishery in 2018

The reported landings in 2018 were $235.9 t$, despite the TAC of $695 t$ only increased slightly from 2017 (Table 7.1). Since 2009, Irish landings of cod reported from ICES rectangles immediately north of the IrishSea/CelticSea boundary (ICES rectangles 33E2 and 33E3) havebeen reallocated into the Celtic Sea as they represent a combination of inaccurate area reporting and catches of cod considered by ICES to be part of the Celtic Sea stock (ICES, 2009). The amount of Irish landings transferred from 7a to 7e-k by year is shown below. Total landings for this stock in 2018 were 214.9 t after this re-allocation.

| Year | Tonnes |
| :--- | :--- | :--- |
| 2004 | 108 |
| 2005 | 54 |
| 2006 | 103 |
| 2007 | 527 |
| 2008 | 558 |
| 2009 | 193 |
| 2010 | 143 |
| 2011 | 147 |
| 2012 | 130 |
| 2013 | 75 |
| 2014 | 24 |
| 2018 | 39 |
|  | 40 |
|  | 19 |
|  | 20 |

The total quota uptake was less than the TAC advice for all nations. Landings by UK vessels have realised $64.3 \%$ of TAC in 2018 (Table 7.2), while the uptake of the TAC allocation in Ireland was low at $18.4 \%$ (Table 7.2). The majority of landings was taken by the TR1 fleet, followed by bycatch in Nephrops trawlers. Landings and discards by métier and country can be seen in Table 7.3. Total uptake of cod TAC was $31 \%$.

A Fishery-Science Partnership Survey (FSP) was repeated in the western Irish Sea in spring 2018 in the western Irish Sea using semi-pelagic gear on commercial vessels. This survey attempts to address the lack of sampling opportunities created by the diminishing TAC for cod in the Irish Sea and the resulting significant reduction of a directed whitefish fleet targeting cod.

All sources of information on age composition in the stock, from the fishery as well as surveys using research vessels and chartered commercial vessels, started to showan increase of cod older than three years of age in the Irish Sea. However, in 2017 those fish turned up in catches at a much lower proportion than expected. Historically the proportion-at-age from the data collected during the sentinel fisheries supports a very steep age profile.

## InterCatch procedure

Since 2013 international landings and discards-at-age are uploaded into InterCatch. Discards are raised for unreported strata and métiers to estimate total discards-at-age.

## Landings

The input data on fishery landings and age compositions are split into four periods:

1) 1968-1990. Landings in this period, provided to ICES by stock coordinators from all countries, are assumed tobeun-biased and are used directly as the input data to stock assessments.

2 ) 1991-1999.TAC reductions in this period caused substantial misreporting of cod landings into several major ports in one country, mainly species misreporting. Landings into these ports were estimated based on observations of cod landings by different fleet sectors during regular port visits. For other national landings, the WG figures provided to ICES stock coordinators were used.
3 ) 2000-2005. Cod recovery measures were considered to have caused significant problems with estimation of landings. The ICES WG landings data provided by stock coordinators for all countries are considered uncertain and estimated within an assessment model. Observations of misreported landings were available for 2000,2001,2002 and 2005. However, they have generally not been used to correct the reported landings but have been used to evaluate model estimates in those years.
4 ) 2006-2018. The introduction of the UK buyers and sellers legislation is considered to have reduced the bias in the landings data but the level to which this has occurred is unknown. Consequently comparisons were made between the fit of the model to recorded landings under an assumption of bias and unbiased information.
The annual numbers-at-age caught and the mean weights-at-age in landings (applied to the total catch) by age are given in Tables 7.2.4 and 7.2.5 and Figures 7.1 and 7.3. Weights-at-age prior to 1982 are fixed at constant values lower than estimated for subsequent years, leading to sum-ofproducts errors, and weights-at-ages $6+$ are becoming noisy for the last few years (Figure 7.3).

Recent years' surveys and commercial data show an improvement in age structure, which resulted from very low fishing pressure since 2013 and a relatively strong 2013 cohort. However this particular cohort seems to have largely disappeared in 2017 and the very poor recruitment in 2016 results in lower landings than expected in 2018.

## Discards data

Discard data (Table $7.6 \mathrm{a}-\mathrm{b}$ ) have been included in the analytical assessments. Landings and discards are combined to catch weight and numbers.

The Cod 7.a Stock Annex and WKIrish3 (ICES, 2017 a, b) benchmark report gives details on historic raising to total national and international discards.

## Biological data

## Natural mortality

Natural mortality has been revised in WKIrish2 (ICES, 2016). M-at-age calculated following Lorenzen (1996) was considered a better representation of the natural mortality than $\mathrm{M}=0.2$. Natural mortality was kept constant throughout years.

## Maturity

Maturity ogive has been revised in WKIrish2 (ICES, 2016). Each year the smoother is applied to the full time-series of raw data and values are accordingly updated. Updated values after application of the smoother are in Table 7.7. Please refer to the stock annex for further information.

## Survey data used in assessment

Please refer to the stock annex for a description of the surveys and survey data. For the current assessment, data for all four surveys were available (Table 7.8).

| Survey | Ages | Years |  |
| :--- | :---: | :---: | :---: |
| NIGFS-WIBTS-Q1 | $1-4$ | 1993 | 2018 |
| NIGFS-WIBTS-Q4 | $0-2$ | 1993 | 2018 |
| UK-FSPw | $2-6$ | 2005 | 2018 (except 2014) |
| NIMIK | 0 | 1994 | 2018 |

## Internal consistency of survey data

The survey data during spring each year are of critical importance for the fit of the assessment models as noted by WGCSE previously and evaluated by WKIrish3 (ICES, 2017a). The data for all surveys were screened by WKIrish3, and due to the number of plots produced, only few are presented here in Figures 7.4-7.7.

## Commercial cpue

Commercial cpue data are available for this stock but are not currently used in the assessment.

### 5.2 Historical stock development

Analytical Model used: ASAP
Due to a bias in the retrospective using 5 peels of $92 \%$ and $52 \%$ for SSB and F respectively, the working group decided that an advice based on this assessment is not reliable. A number of
options were investigated to improve retrospective bias, but eventually the working group concluded to change the stock category of Irish Sea cod from category 1 to category 3. The advice of the stock is now based on a trend-based assessment using a relative biomass and harvest index.

All fits for the ASAP model are supplied at the end of the chapter.

## Deviations from Stock Annex

The assessment did not follow the stock annex as the model provided too great a bias. A trend based assessment was used. The analytical ASAP model is represented here nonetheless as a general trend and background for the change in category.

## Software used and model options chosen

Input data types and characteristics
New data added to the ASAP assessment are the fishery catch data and survey data for 2018. Maturity ogive smoother was applied to the full NIGFS-Q1 Index maturity time-series data to produce a new maturity ogive.

## ASAP model Diagnostics

The diagnostics of the update ASAP run are given in Figure 7.8-7.17.
Figure 7.8 shows the fit of observed and predicted total catches.
Figure 7.9 presents the fitted catch-at-age data for the commercial fleet and the residuals of the fit of the time-series model to the catch data for each age. The fitted values track the trends in the observations well in the early years in which there is no calibration information, with no strong pattern in the residuals. After the introduction of the tuning data, the residuals are increasingly noisy especially for age classes 1 and 2 .

The diagnostics for the Indices are presented in Figures 7.10-7.12.
Figure 7.13 presents the selectivity-at-age of the fishery in two selectivity blocks. The first selectivity block represents the fishery until 1999, the dome-shaped selectivity curve from 2000, as described in the stock annex (ICES, 2107b).

Figure 7.14 presents the estimated selectivity parameters at-age for the time-series of the surveys used in the assessment.

Figure 7.16 shows the fit of the model RMSE. The fit to the total catch and MIKNET index is not perfect, however it provides a reasonable fit to the other surveys.

Retrospective summary for 2013-2018 is displayed in Figure 7.17. Fbar and SSB have been readjusted annually; $\mathrm{Fbar}_{\text {i }}$ is revised upwards, while SSB is revised downwards considerably.

## Final assessment

A trends based analysis was eventually used to assess the stock status.
The NIGFS-Q1 survey was used to generate a biomass trend by multiplying the relative abun-dance-at-age/nautical miles (Table 7.8 a) by the weight-at-age (Table 7.5b) and summing within each year. This was then standardised to the mean and the average of the last two years was divided by the mean of the preceding three to generate the index of change. The rate of change (resulting from the 2 over 3 rule) (Figure 7.18, Table 7.9) was 0.57 , suggesting that the biomass has been largely stable over the past few years and had a considerate drop in 2018.

The biomass index only takes into account ages $1-4$, excluding the larger fish which have been responsible for the apparent increase in biomass in the ASAP assessment.

A harvest rate indicator was constructed using the ratio of biomass indicator to total catches which was then standardised to the mean.

## Final assessment: long-term trends

The harvest rate indicator has been constantly declining since 2013 and only sawa slight increase in 2018 due to the landings increase in response to the higher TAC (Figure 7.18, Table 7.9) and at the same time a decline in the biomass index. The same general decline is observed in the ASAP assessment.

The biomass index shows a mostly stable biomass trend in contrast to the (biased) ASAP assessment which shows an increasing trend in SSB in recent years. The biomass trend takes into account fish from ages 1 to 4 , of which not all constitute of the SSB, but largely follows the commercial exploitation pattern and is therefore considered an appropriate index for exploitable biomass.

## The state of the stock from ASAP model

Spawning-stockbiomasshas declined ten-fold since the late 1980s and recruitment has beenlow since the mid-1990s, particularly since 2000. Fishing mortality has been declining in recent years and has dropped to below FMSY. Since 2010 SSB has slowly recovered (Figure 7.17).

Fishing mortality throughout large parts of the assessment period has been well above the candidate reference points associated with high long-term yields and a low risk of depleting the productive potential of the stock. The assessment shows a steep decline in F from 2012 (Figure 7.17).

Recruitment has been low for the past eighteen years. 2012 and 2013 year classes have increased recruitment, but were still well below the long-term average. Recruitment in 2016 was estimated at the historically lowest point (Figure 7.17).

The status of the stock in future years is very uncertain.

## Investigated options of the analytical model

A range of options were discussed during the working group meeting to decrease the retrospective bias in the model. While some of the options looked promising, further work will have to be done.

The options included:

- Changing the index selectivity pattern of Q1 survey;
- Keeping M for older ages at the level of first maturity;
- Adding a third selectivity block to commercial catches;
- Removing the final year of the FSP index, as this might have been caused by a year effect;
- Removing the Q4 survey index, since it is not tracking cohorts to a high degree;
- Curtailing the age structure of cod to age 4 (age $4+$ group).


### 5.3 Short-term predictions

Due to the stock being re-classified as an ICES category 3 stock, no short-term forecast was carried out.

### 5.4 Biological reference points

No reference points available at the current time, as stock has been re-classified as category 3 .

### 5.5 Management plans

The Irish Sea cod management plan, as described in Council Regulation(EC) 1342/2008 was evaluated independently by ICES in 2009 using the approach adopted in AGCREMP 2008 and found to be not consistent with the ICES Precautionary Approach (WGCSE 2009).

### 5.6 Uncertainties and bias in assessment

## Landings data

The quality of the commercial landings and catch-at-age data for this stock deteriorated in the 1990s following reductions in the TAC without associated control of fishing effort. The Working Group has, since the 1990s, attempted to overcome this problem by incorporating sample-based estimates of landings from three major ports in the WG landings figures. The data for this method have been poor for the years 2003-2006, hence data for this period have been estimated by the WG using modelling approaches.

## Discarding

Discarding has historically been mainly at-age 0 and 1 .
The Irish Sea whitefish fleet has got good observer coverage as does the Nephrops fleet except for the years 2003-2006.

Strict controls on landings reporting following the introduction of the Registration of Fish Buyers and Sellers regulations has resulted in documented increases in discarding of older cod in the Irish Sea since 2012 (Figure 7.2).

Compliance with catch composition rules for some fleets, especially for those targeting Nephrops, could also result in increased discarding of cod.

The increased TAC for cod in the Irish Sea in 2018 will hopefully lead to lower discard rates of older cod in the Nephrops fleet.

## Surveys

The Irish Sea has relatively good survey coverage. The surveys in general give consistent signals of fish abundance-at-age (Figure 7.6).

The UK Fisheries Science Partnership surveys (UK-FSP) of the Irish Sea cod spawning grounds in spring 2005-2018 carried out using commercial trawlers, indicated a widespread distribution of cod mostly at low densitybut with some localized aggregations. The time-series of SSB indices shows an upward trend similar to that shown by NIGFS-WIBTS-Q1 pointing to some recovery following the maturation of the 2012 and 2013 year classes, however there was a considerable drop in 2018. The drop resulted from the very low 2016 year class entering the fishery in 2018.

## Stock structure and migrations

Stock structure and migrations have been in full discussed in the WKIrish2 report (ICES, 2016).
A tagging study of Irish Sea cod began in 2016 in part to address these issues. Up to January 2019 4238 cod were caught and tagged aboard chartered commercial fishing vessel using semi-pelagic
fishing gear, FSP survey, shore angling competitions and others. Up to January 2019, 138 tagged cod were returned. The project relies on collaboration with the fishing industry to provide the data to develop a better understanding of the current behaviour, biology and stock status of Irish Sea cod. Most recent results suggest a stronger migratory behaviour of Irish Sea cod into the Celtic Sea, indicating that up to $18 \%$ of mature fish might leave the Irish Sea. This will have considerable impacts on the future management and assessment of the stock, but additional research is necessary.

### 5.7 Management considerations

A number of emergency and cod recovery plan measures have been introduced since 2000 to conserve Irish Sea cod. These include a spawning closure since 2000 and effort control since 2003. There have also been several vessel decommissioning schemes. As it has not been possible to provide analytical catch forecasts in recent years, the TAChas been reduced by 15-20\% annually since 2006 and by $25 \%$ since 2009. In 2017 all sources of information on age composition in the stock, from the fishery as well as surveys using research vessels and chartered commercial vessels, indicated a trend towards a recovery of the stock and a decline in fishing pressure.

### 5.8 Future Issues and considerations

Cod in the IrishSea and the Celtic Sea are in a highly exploited state and show historically a very steep age-profile. With the decline of the active fishery since 2000, there has been slight improvements in the stocks, however it is questionable in how far traditional analytical assessment methodologies are able to be used in the assessment.

The recent years show that a single, above average, cohort (the 2013 year class), can have a considerable impact on the SSB. However, as those fish seemed to disappear at-age 3 or 4 , this resulted in the strong retrospective downwards revisions of SSB.

At the current state it is unclear as to what happened to the fish once they reached age 3 to 4 , and the analytical model keeps revising the size of the year class down to fit.

It is essential to further understanding of the stock structure to improve future management, which includes the further investigation of migration and natural mortality in the Irish Sea. It might be necessary for a combined approach to manage the stocks in 7A and 7E-G, in the light of the recent tagging study results.

### 5.9 References

Armstrong, M. J., Gerritsen, H. D., Allen, M., McCurdy, W. J. and Peel, J. A. D. 2004. Variability in maturity and growth in a heavily exploited stock: cod (Gadus morhua L.) in the Irish Sea. ICES J. Mar. Sci., 61, 98-112.
Bendall, V. O., Ó'Cuaig, M., Schon, P-J., Hetherington, S., Armstrong, M., Graham, N., Righton, D. 2009. Spatio-temporal dynamics of Atlantic cod (Gadus morhua) in the Irish and Celtic Sea: results from a collaborative tagging programme. ICES Document CM 2009/J: 06.35 pp .

ICES. 2016. Report of the Data Evaluation Workshop on Irish Sea Fisheries (WKIrish2), 26-29 September 2016, Belfast, UK, ICES CM 2016/BSG:02.

ICES. 2017a. Report of the Benchmark Workshop on Irish Sea fish (WKIrish3), 30 January-3 February 2017, Galway, ROI, ICES CM 2017/BSG:01.

Lorenzen, K. 1996. The relationship between body weight and natural mortality in fish: a comparison of natural ecosystems and aquaculture. J. Fish Biol., 49: 627-647.
STECF. 2011. Scientific, Technical and Economic Committee for Fisheries. Evaluation of Fishing Effort Regimes Regarding Annexes IIA, IIB and IIC of TAC \& Quota Regulations, Celtic Sea and Bay of Biscay (STECF-11-13).

STECF. 2013. Evaluation of Fishing Effort Regimes in European Waters - Part 2 (STECF-13-21). Publications Office of the European Union, Luxembourg, EUR 26327 EN, JRC86088, 863 pp.

Table 7.1. Nominal landings ( $\mathbf{t}$ ) of COD in Division 7.a as officially reported to ICES and figures used by ICES from 1999.

| Country | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | $2017{ }^{1}$ | $2018{ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 150 | 60 | 283 | 318 | 183 | 104 | 115 | 60 | 67 | 26 | 19 | 21 | 36 | 23 | 13 | 9 | 12 | 3 | 5 | 1.9 |
| France | $\mathrm{n} / \mathrm{a}$ | 53 | 74 | 116 | 151 | 29 | 35 | $18^{2}$ | $17^{2}$ | 3 | 12 | 1 | 3 | 1 | <1 | <1 | <1 | <1 | <1 | <1 |
| Ireland | 966 | 455 | 751 | 1,111 | 594 | 380 | 220 | 275 | 608 | $618^{2}$ | $323^{2}$ | 289 | 275 | 193 | 160 | 148 | 137 | 84.2 | 57.2 | 104.6 |
| Netherlands | 5 | 1 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Spain | - | - | - | - | 14 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| UK (England, Wales \& NI) | 1,665 | 799 | 885 | 1,134 | 505 | 646 | 594 | 5892 | 423 | 5432 | 3872 | 282 | 169 | 109 | 107 | 79 | 50 | 35.5 | 41.1 | 113.3 |
| UK (Isle of Man) | 9 | 11 | 1 | 7 | 7 | 5 | n/a | n/a | n/a | 22 | 12 | 1 | 1 | <1 | <1 | <1 | <1 | <1 | <1 | <1 |
| UK (Scotland) | 80 | 38 | 32 | 29 | 23 | 15 | 3 | 6 | 2 | 12 | 12 | - | - | - | - | - | - | - | <1 | <1 |
| Total | 2,875 | 1,417 | 2,026 | 2,715 | 1,477 | 1,179 | 967 | 948 | 1,117 | 1224 | 754 | 594 | 485 | 326 | 281 | 236 | 199 | 122.83 | 103.85 | 234.9 |
| Unallocated | 1,909 | -143 | 226 | -20 | -192 | -107 | -57 | -108 | -415 | -563 | -286 | -130 | -117 | -128 | -75 | -33 | -38 | -40.5 | -19.4 | -20 |
| Total as used by WG | $4784^{3}$ | $1274{ }^{4}$ | $2252{ }^{4}$ | $2695{ }^{4}$ | $1285{ }^{4}$ | $1072^{4}$ | $910^{4}$ | $840^{4}$ | 7024 | $661{ }^{4}$ | $468{ }^{4}$ | $464{ }^{4}$ | 368 | 198 | 206 | 213 | 161 | 82 | 84 | 214.9 |

${ }^{1}$ Preliminary. ${ }^{2}$ Revised. $\mathrm{n} / \mathbf{a}=$ not available ${ }^{3}$ includes sample-based estimates of landings into three ports ${ }^{4}$ based on official data only.

Table 7.2. a)-c) Cod in 7a. Working Group figures for annual landings and TAC uptake by country since 2000 (2009).
a)

| Year | NI | E \& W | Scotland | Ireland | France | Belgium | Isle of Man | Netherlands | Total | TAC | \% uptake |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 638 | 156 | 39 | 321 | 52 | 56 | 11 | 0 | 1273 | 2100 | 61 |
| 2001 | 697 | 209 | 32 | 645 | 361 | 300 | 8 | 0 | 2251 | 2100 | 107 |
| 2002 | 983 | 171 | 39 | 953 | 251 | 294 | 1 | 2 | 2695 | 3200 | 84 |
| 2003 | 381 | 118 | 32 | 415 | 145 | 187 | 7 | 0 | 1285 | 1950 | 66 |
| 2004 | 539 | 103 | 15 | 271 | 37 | 103 | 5 | 0 | 1072 | 2150 | 50 |
| 2005 | 523 | 72 | 4 | 168 | 31 | 108 | 3 | 0 | 910 | 2150 | 42 |
| 2006 | 552 | 32 | 6 | 172 | 17 | 59 | 3 | 0 | 840 | 1828 | 46 |
| 2007 | 396 | 27 | 2 | 191 | 18 | 66 | 2 | 0 | 702 | 1462 | 48 |
| 2008 | 523 | 22 | 1 | 85 | 3 | 27 | 1 | 0 | 662 | 1199 | 55 |
| 2009 | 375 | 15 | 0 | 55 | 3 | 19 | 1 | 0 | 468 | 899 | 52 |
| 2010 | 274 | 17 | 0 | 151 | 1 | 21 | 1 | 0 | 465 | 674 | 69 |
| 2011 | 152 | 17 | 0 | 160 | 3 | 36 | 1 | 0 | 368 | 506 | 73 |
| 2012 | 98 | 14 | 0 | 63 | 0 | 23 | 0 | 0 | 198 | 380 | 52 |
| 2013 | 103 | 4 | 0 | 85 | 1 | 13 | 0 | 0 | 206 | 285 | 72 |
| 2014 | 72 | 7 | 0 | 124 | 0 | 9 | 0 | 0 | 213 | 182 | 117 |
| 2015 | 47 | 3 | 0 | 99 | 0 | 12 | 0 | 0 | 161 | 146 | 110 |
| 2016 | 32 | 3 | 0 | 45 | 0.4 | 3 | 0 | 0 | 82 | 146 | 56 |
| 2017 | 38 | 3 | 0.2 | 38 | 0.2 | 5 | 0.05 | 0 | 84 | 146 | 57 |
| 2018 | 113 | 15 | <0.1 | 85 | <0.1 | 2 | $<0.1$ | 0 | 215 | 695 | 31 |

b)

| 2009 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 391 | 55 | 3 | 19 | 0 | 498 |
| TAC | 259 | 592 | 33 | 12 | 3 | 899 |
| $\%$ uptake | $151 \%$ | $9 \%$ | $9 \%$ | $160 \%$ | $0 \%$ |  |


| 2010 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 292 | 151 | 1 | 21 | 0 | 465 |
| TAC | 194 | 444 | 25 | 9 | 2 | 674 |
| \% uptake | $150 \%$ | $34 \%$ | $4 \%$ | $233 \%$ | $0 \%$ |  |


| 2011 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 170 | 160 | 3 | 36 | 0 | 369 |
| TAC | 146 | 333 | 19 | 7 | 2 | 506 |
| \% uptake | $117 \%$ | $48 \%$ | $16 \%$ | $533 \%$ | $0 \%$ |  |


| 2012 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 112 | 63 | 0 | 23 | 0 | 198 |
| TAC | 109 | 251 | 14 | 5 | 1 | 380 |
| \% uptake | $103 \%$ | $25 \%$ | $0 \%$ | $460 \%$ | $0 \%$ |  |


| 2013 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 107 | 85 | 1 | 13 | 0 | 206 |
| TAC | 82 | 188 | 10 | 4 | 1 | 285 |
| $\%$ uptake | $130 \%$ | $45 \%$ | $10 \%$ | $325 \%$ | $0 \%$ |  |


| 2014 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 79 | 124 | 0 | 9 | 0 | 213 |
| TAC | 52 | 120 | 7 | 2 | 2 | 182 |
| \% uptake | $153 \%$ | $103 \%$ | $0 \%$ | $455 \%$ | $0 \%$ |  |


| 2015 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 50 | 99 | 0 | 12 | 0 | 161 |
| TAC | 42 | 97 | 5 | 2 | 0 | 146 |
| \% uptake | $119 \%$ | $102 \%$ | $0 \%$ | $600 \%$ | NA |  |


| 2016 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | ---: | :---: | ---: | :---: | :---: | :---: |
| Landings | 35 | 44 | 0.4 | 3 | 0 | 82 |
| TAC | 42 | 97 | 5 | 2 | 0 | 146 |
| $\%$ uptake | $83 \%$ | $45 \%$ | $8 \%$ | $150 \%$ | $0 \%$ |  |
|  |  |  |  |  | 0 | Total |
| UK | Ireland | France | Belgium | Netherlands | 84 |  |
| Landings | 41 | 38 | 0.2 | 5 | 0 | 146 |
| TAC | 42 | 97 | 5 | 2 | $0 \%$ |  |
| $\%$ uptake | $98 \%$ | $39 \%$ | $4 \%$ | $250 \%$ | 0 |  |


| 2018 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 128.5 | 84.6 | 0.05 | 1.9 | 0 | 214.9 |
| TAC | 200 | 459 | 25 | 9 | 2 | 695 |
| $\%$ uptake | $64 \%$ | $18 \%$ | $<1 \%$ | $<1 \%$ | $0 \%$ | $31 \%$ |

c) Landings proportions by country since 2000 .

| Year | NI | E \& W | Scotland | Ireland | France | Belgium | Isle of Man | Netherlands | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 50.1 | 12.3 | 3.0 | 25.2 | 4.1 | 4.4 | 0.9 | 0.0 | 100 |
| 2001 | 31.0 | 9.3 | 1.4 | 28.6 | 16.1 | 13.3 | 0.4 | 0.0 | 100 |
| 2002 | 36.5 | 6.4 | 1.5 | 35.4 | 9.3 | 10.9 | 0.0 | 0.1 | 100 |
| 2003 | 29.7 | 9.2 | 2.5 | 32.3 | 11.3 | 14.6 | 0.6 | 0.0 | 100 |
| 2004 | 50.3 | 9.6 | 1.4 | 25.2 | 3.5 | 9.6 | 0.4 | 0.0 | 100 |
| 2005 | 57.5 | 7.9 | 0.5 | 18.5 | 3.5 | 11.8 | 0.3 | 0.0 | 100 |
| 2006 | 65.7 | 3.8 | 0.7 | 20.4 | 2.0 | 7.1 | 0.3 | 0.0 | 100 |
| 2007 | 56.5 | 3.8 | 0.3 | 27.2 | 2.5 | 9.5 | 0.3 | 0.0 | 100 |
| 2008 | 78.9 | 3.4 | 0.2 | 12.8 | 0.5 | 4.0 | 0.2 | 0.0 | 100 |
| 2009 | 80.1 | 3.1 | 0.0 | 11.7 | 0.6 | 4.1 | 0.3 | 0.0 | 100 |
| 2010 | 41.3 | 4.6 | 0.0 | 43.5 | 0.8 | 9.8 | 0.2 | 0.0 | 100 |
| 2011 | 41.3 | 4.6 | 0.0 | 43.5 | 0.8 | 9.8 | 0.3 | 0.0 | 100 |
| 2015 | 49.5 | 7.1 | 0.0 | 31.8 | 0.0 | 11.6 | 0.0 | 0.0 | 100 |
| 2013 | 50.0 | 1.9 | 0.1 | 41.3 | 0.2 | 6.3 | 0.2 | 0.0 | 100 |
| 2014 | 33.8 | 3.3 | 0.0 | 58.2 | 0.0 | 4.2 | 0.0 | 0.0 | 100 |
| 2015 | 29.2 | 1.9 | 0.0 | 61.5 | 0.0 | 7.5 | 0.0 | 0.0 | 100 |
| 2016 | 39.0 | 3.7 | 0.0 | 54.9 | 0.5 | 3.7 | 0.0 | 0.0 | 100 |
| 2017 | 45.5 | 3.2 | 0.3 | 44.7 | 0.2 | 6 | 0.1 | 0.0 | 100 |
| 2018 | 52.7 | 7 | <0.1 | 39.4 | 0.0 | 0.9 | <0.1 | 0.0 | 100 |

Table 7.3. Landings and discard proportions by métier.

| Catch (2018) | Estimated landings |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $257$ <br> tonnes | otter trawls |  | Scottish seines | mid-water trawl | beam trawls | other gear types |
|  | Nephrops directed | demersal fish directed | <1\% | 16.8\% | 9.1\% | 1.15\% |
|  | 25\% | 48\% |  |  |  |  |
|  | 215 tonnes |  |  |  |  |  |
|  | Estimated discards |  |  |  |  |  |
|  | otter trawls |  | Scottish seines | mid-water trawl | beam trawls | other gear types |
|  | 70\% Nephrops directed | 8.3\% demersal fish directed | <1\% | <1\% | 19.5\% | 1.85\% |
|  | 42 tonnes |  |  |  |  |  |

Table 7.4. Cod in 7a. Total catch numbers-at-age (thousands).

|  |  | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 17 | 439 | 1563 | 1003 | 456 | 177 | 30 |
| 1969 | 20 | 969 | 1481 | 1050 | 269 | 186 | 113 |
| 1970 | 22 | 1810 | 1385 | 352 | 204 | 163 | 71 |
| 1971 | 22 | 2835 | 2022 | 904 | 144 | 67 | 51 |
| 1972 | 26 | 900 | 3267 | 824 | 250 | 58 | 59 |
| 1973 | 27 | 2377 | 1091 | 1783 | 430 | 173 | 81 |
| 1974 | 16 | 601 | 3559 | 557 | 494 | 131 | 74 |
| 1975 | 26 | 1810 | 642 | 1407 | 294 | 249 | 117 |
| 1976 | 27 | 1247 | 3007 | 363 | 500 | 61 | 104 |
| 1977 | 31 | 946 | 511 | 1233 | 163 | 218 | 71 |
| 1978 | 40 | 855 | 1092 | 310 | 311 | 39 | 65 |
| 1979 | 44 | 1948 | 1288 | 608 | 127 | 164 | 71 |
| 1980 | 25 | 2636 | 2797 | 729 | 243 | 49 | 55 |
| 1981 | 38 | 1457 | 3635 | 1448 | 244 | 99 | 47 |
| 1982 | 46 | 538 | 2284 | 1455 | 557 | 102 | 79 |
| 1983 | 47 | 1011 | 932 | 751 | 499 | 154 | 46 |
| 1984 | 37 | 1733 | 1195 | 439 | 240 | 161 | 75 |
| 1985 | 34 | 1360 | 2105 | 703 | 158 | 84 | 77 |
| 1986 | 49 | 1180 | 2248 | 699 | 203 | 64 | 65 |
| 1987 | 47 | 4522 | 1793 | 841 | 252 | 75 | 43 |
| 1988 | 43 | 2971 | 4734 | 702 | 263 | 71 | 38 |


|  |  | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 41 | 754 | 2163 | 1886 | 231 | 86 | 37 |
| 1990 | 38 | 869 | 1075 | 545 | 372 | 70 | 30 |
| 1991 | 47 | 2169 | 1408 | 442 | 127 | 98 | 22 |
| 1992 | 37 | 1529 | 1243 | 664 | 132 | 42 | 49 |
| 1993 | 39 | 388 | 2907 | 403 | 119 | 16 | 13 |
| 1994 | 40 | 916 | 569 | 848 | 68 | 20 | 10 |
| 1995 | 43 | 678 | 1283 | 180 | 163 | 7 | 6 |
| 1996 | 88 | 447 | 1113 | 700 | 38 | 39 | 6 |
| 1997 | 5 | 651 | 1149.5 | 501 | 213 | 17 | 16 |
| 1998 | 0 | 231 | 1928 | 335 | 80 | 28 | 8 |
| 1999 | 141 | 236 | 843 | 871 | 66 | 21 | 7 |
| 2000 | 62 | 1107 | 176 | 107 | 50 | 4 | 1 |
| 2001 | 7 | 403 | 841 | 53 | 13 | 9 | 2 |
| 2002 | 0 | 238 | 564 | 405 | 7 | 2 | 3 |
| 2003 | 50 | 121 | 472 | 109 | 36 | 1 | 0 |
| 2004 | 50 | 161 | 134 | 174 | 22 | 6 | 3 |
| 2005 | 50 | 118 | 256 | 78 | 34 | 5 | 1 |
| 2006 | 50 | 89 | 174 | 128 | 17 | 8 | 3 |
| 2007 | 16 | 216 | 210 | 56 | 11 | 1 | 0 |
| 2008 | 6 | 77 | 169 | 87 | 9 | 3 | 0 |
| 2009 | 329 | 60 | 57 | 66 | 17 | 3 | 0 |
| 2010 | 49 | 220 | 188 | 16 | 7.5 | 2 | 1 |
| 2011 | 10 | 54 | 106 | 36 | 2 | 1 | 1 |
| 2012 | 8 | 84 | 135 | 145 | 10 | 0 | 0 |
| 2013 | 36 | 37 | 59 | 30 | 9 | 2 | 0 |
| 2014 | 1 | 41 | 86 | 26 | 5 | 1 | 0 |
| 2015 | 0 | 37 | 80 | 26 | 4 | 1 | 0 |
| 2016 | 0 | 11 | 25 | 30 | 2 | 1 | 0 |
| 2017 | 0 | 12 | 28 | 16 | 3 | 0 | 0 |
| 2018 | 256 | 95 | 27 | 36 | 2 | 2 | 1 |

Table 7.5. a) Mean weights-at-age in the landings (used for whole stock and catch).

|  |  | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.86 |
| 1969 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.26 |
| 1970 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.17 |
| 1971 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.12 |
| 1972 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.28 |
| 1973 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.16 |
| 1974 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.34 |
| 1975 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.05 |
| 1976 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.13 |
| 1977 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.63 |
| 1978 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.19 |
| 1979 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.48 |
| 1980 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.87 |
| 1981 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.55 |
| 1982 | 0.1 | 1.01 | 1.52 | 3.49 | 5.57 | 7.59 | 9.11 |
| 1983 | 0.1 | 1 | 1.84 | 3.99 | 5.96 | 7.97 | 9.97 |
| 1984 | 0.1 | 0.68 | 1.81 | 3.81 | 5.87 | 7.48 | 10.05 |
| 1985 | 0.1 | 0.78 | 2.02 | 4.24 | 5.83 | 7.5 | 9.04 |
| 1986 | 0.1 | 0.81 | 1.83 | 3.86 | 5.86 | 7.39 | 8.78 |
| 1987 | 0.1 | 0.71 | 2.16 | 3.91 | 6.41 | 7.82 | 10.32 |
| 1988 | 0.1 | 0.61 | 1.56 | 3.76 | 5.67 | 8.02 | 9.88 |
| 1989 | 0.1 | 0.94 | 1.85 | 3.22 | 5.41 | 6.57 | 9.47 |
| 1990 | 0.1 | 0.84 | 1.94 | 3.57 | 5.28 | 7.53 | 9.4 |
| 1991 | 0.1 | 0.86 | 1.64 | 3.54 | 5.42 | 6.39 | 9.11 |
| 1992 | 0.1 | 0.81 | 1.96 | 3.99 | 5.98 | 6.92 | 8.67 |
| 1993 | 0.1 | 0.85 | 1.71 | 3.67 | 5.68 | 7.37 | 10.17 |
| 1994 | 0.1 | 0.8 | 1.92 | 3.61 | 6.08 | 7.68 | 8.57 |
| 1995 | 0.1 | 0.9 | 1.84 | 4.00 | 5.79 | 8.45 | 9.14 |


|  |  | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 0.1 | 0.98 | 1.63 | 3.26 | 5.3 | 7.72 | 9.79 |
| 1997 | 0.1 | 0.85 | 1.94 | 3.62 | 5.29 | 6.12 | 9.4 |
| 1998 | 0.1 | 0.93 | 1.65 | 3.73 | 5.37 | 7.03 | 9.35 |
| 1999 | 0.1 | 0.85 | 1.62 | 3.18 | 5.51 | 7.52 | 10.25 |
| 2000 | 0.1 | 0.85 | 1.99 | 3.57 | 5.14 | 7.15 | 8.39 |
| 2001 | 0.1 | 0.99 | 1.82 | 4.15 | 5.61 | 7.33 | 9.51 |
| 2002 | 0.1 | 0.94 | 1.84 | 3.44 | 5.73 | 7.71 | 10.01 |
| 2003 | 0.1 | 1.21 | 1.66 | 3.29 | 5.43 | 10.2 | 11.09 |
| 2004 | 0.1 | 1.11 | 2.2 | 3.63 | 6.51 | 7.64 | 8.61 |
| 2005 | 0.1 | 0.91 | 1.94 | 3.51 | 5.32 | 7.74 | 8.89 |
| 2006 | 0.1 | 0.83 | 1.84 | 3.67 | 4.71 | 6.39 | 7.84 |
| 2007 | 0.1 | 0.83 | 1.85 | 3.78 | 5.35 | 7.99 | 10.04 |
| 2008 | 0.1 | 0.89 | 1.59 | 3.54 | 6.00 | 7.57 | 9.46 |
| 2009 | 0.1 | 1.1 | 2.01 | 3.46 | 5.31 | 7.1 | 6.82 |
| 2010 | 0.1 | 1.26 | 2.29 | 3.93 | 6.34 | 7.33 | 9.64 |
| 2011 | 0.1 | 0.95 | 1.88 | 3.75 | 5.54 | 6.75 | 9.04 |
| 2012 | 0.1 | 0.93 | 1.88 | 3.37 | 5.34 | 7.60 | 8.56 |
| 2013 | 0.1 | 0.97 | 2.32 | 4.06 | 5.54 | 7.43 | 10.79 |
| 2014 | 0.1 | 0.88 | 2.26 | 4.49 | 7.00 | 8.75 | 9.41 |
| 2015 | 0.1 | 0.83 | 1.79 | 3.69 | 6.49 | 8.55 | 9.95 |
| 2016 | 0.1 | 0.95 | 1.58 | 3.1 | 5.01 | 10.66 | 8.136 |
| 2017 | 0.1 | 0.70 | 1.82 | 3.82 | 5.85 | 7.62 | 9.74 |
| 2018 | 0.1 | 0.43 | 1.69 | 3.64 | 5.56 | 8.58 | 8.70 |

b) Q1 survey weights used to calculate the biomass index.

| Year | 1 | 2 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: |
| 1993 | 0.18 | 0.97 | 3.01 | 5.12 |
| 1994 | 0.14 | 1.25 | 3.06 | 4.82 |
| 1995 | 0.11 | 0.92 | 3.21 | 5.27 |
| 1996 | 0.17 | 1.08 | 2.65 | 7.20 |
| 1997 | 0.13 | 1.21 | 3.24 | 4.65 |
| 1998 | 0.19 | 0.91 | 4.01 | 3.96 |
| 1999 | 0.22 | 1.04 | 2.81 | 3.93 |
| 2000 | 0.17 | 1.55 | 3.69 | 4.85 |
| 2001 | 0.12 | 1.13 | 3.98 | 4.98 |
| 2002 | 0.15 | 1.20 | 3.11 | NA |
| 2003 | 0.11 | 1.02 | 2.77 | 5.04 |
| 2004 | 0.12 | 1.24 | 2.96 | 6.21 |
| 2005 | 0.12 | 1.50 | 3.67 | 5.52 |
| 2006 | 0.10 | 1.04 | 3.05 | 4.95 |
| 2007 | 0.10 | 1.07 | 3.58 | 7.05 |
| 2008 | 0.09 | 1.03 | 2.91 | 4.96 |
| 2009 | 0.18 | 1.26 | 2.82 | 6.50 |
| 2010 | 0.11 | 1.28 | 2.96 | 7.54 |
| 2011 | 0.11 | 0.98 | 3.41 | 4.62 |
| 2012 | 0.14 | 0.98 | 2.35 | 4.74 |
| 2013 | 0.11 | 1.12 | 3.51 | 6.09 |
| 2014 | 0.11 | 1.24 | 3.55 | 6.08 |
| 2015 | 0.10 | 0.91 | 3.52 | 7.74 |
| 2016 | 0.10 | 0.94 | 3.21 | 5.03 |
| 2017 | 0.09 | 0.92 | 3.06 | 5.63 |
| 2018 | 0.18 | 1.17 | 2.56 | NA |

Table 7.6. Cod in 7.a. Estimates of numbers discarded (a) and the discarded proportions (b) from 1968-2018. Data are total numbers (‘000 fish) discarded at-age, estimated from numbers per sampled trip raised to total fishing effort by each country supplying data (UK, Ireland and Belgium) Please refer to WKIrish3 (ICES, 2017a) documents.
a)

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 17.81 | 74.71 | 0 | 0 | 0 | 0 | 0 |
| 1969 | 20.85 | 87.45 | 0 | 0 | 0 | 0 | 0 |
| 1970 | 22.13 | 92.83 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 22.94 | 96.2 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 26.51 | 111.18 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 27.17 | 113.96 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 16.94 | 71.04 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 26.38 | 110.62 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 26.77 | 112.28 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 31.05 | 130.23 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 39.96 | 167.57 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 44.35 | 185.98 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 24.6 | 103.16 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 37.67 | 157.97 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 46.04 | 193.1 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 46.98 | 197.05 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 37.3 | 156.45 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 33.89 | 142.12 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 49.15 | 206.15 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 47.38 | 198.69 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 42.59 | 178.64 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 41.03 | 172.09 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 37.85 | 158.74 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 46.64 | 195.61 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 36.74 | 154.1 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 39.4 | 165.24 | 0 | 0 | 0 | 0 | 0 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 39.92 | 167.44 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 42.97 | 180.2 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 87.95 | 128.79 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 5.28 | 127.79 | 0.5 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 27.47 | 2 | 0 | 0 | 0 | 0 |
| 1999 | 141.42 | 165.79 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 62.36 | 817.69 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 7.22 | 65.15 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 42.49 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 50.43 | 75.68 | 32.62 | 15.83 | 1.25 | 0.13 | 0 |
| 2004 | 50.43 | 92.78 | 32.81 | 15.83 | 1.25 | 0.13 | 0 |
| 2005 | 50.43 | 76.34 | 32.36 | 15.83 | 1.25 | 0.13 | 0 |
| 2006 | 50.43 | 75.08 | 32 | 15.83 | 1.25 | 0.13 | 0 |
| 2007 | 16 | 167 | 4.60 | 0 | 0 | 0 | 0 |
| 2008 | 5.50 | 63.40 | 3.40 | 0 | 0 | 0 | 0 |
| 2009 | 329.30 | 39.80 | 4.40 | 0.1 | 0 | 0 | 0 |
| 2010 | 48.70 | 180 | 60.30 | 1.4 | 0.5 | 0.1 | 0 |
| 2011 | 9.70 | 42.70 | 0.90 | 0 | 0 | 0 | 0 |
| 2012 | 7.50 | 79.90 | 100.20 | 112.9 | 5.9 | 0.2 | 0 |
| 2013 | 36.10 | 31 | 26.50 | 11 | 2 | 0.5 | 0 |
| 2014 | 1.09 | 34.66 | 41.93 | 10.3 | 1.53 | 0.1 | 0 |
| 2015 | 0 | 37.30 | 45.80 | 6.8 | 1.3 | 0.3 | 0 |
| 2016 | 0 | 9.84 | 14.15 | 13.45 | 0.91 | 0.74 | 0 |
| 2017 | 0.43 | 9.85 | 7.88 | 8.10 | 0.57 | 0.10 | 0.10 |
| 2018 | 255.50 | 72.19 | 8.89 | 4.88 | 0.12 | 0.22 | 0 |

b)

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 1 | 0.17 | 0 | 0 | 0 | 0 | 0 |
| 1969 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1970 | 1 | 0.05 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 1 | 0.03 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 1 | 0.12 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 1 | 0.05 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 1 | 0.12 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 1 | 0.06 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 1 | 0.14 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 1 | 0.20 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 1 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 1 | 0.04 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 1 | 0.11 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 1 | 0.36 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 1 | 0.19 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 1 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 1 | 0.17 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 1 | 0.04 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 1 | 0.06 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 1 | 0.23 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 1 | 0.18 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 1 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 1 | 0.43 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 1 | 0.18 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 1 | 0.27 | 0 | 0 | 0 | 0 | 0 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 1 | 0.29 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 1 | 0.20 | 0 | 0 | 0 | 0 | 0 |
| 1998 | NA | 0.12 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 1 | 0.70 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 1 | 0.74 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 1 | 0.16 | 0 | 0 | 0 | 0 | 0 |
| 2002 | NA | 0.18 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 1 | 0.63 | 0.07 | 0.15 | 0.03 | 0.12 | NA |
| 2004 | 1 | 0.58 | 0.25 | 0.09 | 0.06 | 0.022 | 0 |
| 2005 | 1 | 0.65 | 0.13 | 0.20 | 0.04 | 0.03 | 0 |
| 2006 | 1 | 0.84 | 0.18 | 0.12 | 0.07 | 0.02 | 0 |
| 2007 | 1 | 0.77 | 0.02 | 0 | 0 | 0 | NA |
| 2008 | 1 | 0.82 | 0.02 | 0 | 0 | 0 | NA |
| 2009 | 1 | 0.67 | 0.08 | 0 | 0 | 0 | NA |
| 2010 | 1 | 0.82 | 0.32 | 0.06 | 0.07 | 0.05 | 0 |
| 2011 | 1 | 0.80 | 0.01 | 0 | 0 | 0 | 0 |
| 2012 | 1 | 0.95 | 0.74 | 0.78 | 0.60 | 1 | NA |
| 2013 | 1 | 0.84 | 0.45 | 0.37 | 0.22 | 0.34 | NA |
| 2014 | 1 | 0.85 | 0.49 | 0.39 | 0.28 | 0.09 | NA |
| 2015 | NA | 1 | 0.57 | 0.26 | 0.30 | 0.23 | NA |
| 2016 | NA | 0.91 | 0.58 | 0.45 | 0.40 | 0.62 | 0 |
| 2017 | 1 | 0.80 | 0.28 | 0.51 | 0.20 | 0.21 | 0.49 |
| 2018 | 1 | 0.76 | 0.33 | 0.13 | 0.05 | 0.10 | 0 |

NA= not available.

Table 7.7. Maturity ogive updated for 2018. Prior to 1995 maturity was considered constant.

|  | 1 | 2 | 3+ |
| :---: | :---: | :---: | :---: |
| 1996 | 0 | 0.28 | 1 |
| 1997 | 0 | 0.34 | 1 |
| 1998 | 0 | 0.40 | 1 |
| 1999 | 0 | 0.46 | 1 |
| 2000 | 0 | 0.53 | 1 |
| 2001 | 0 | 0.59 | 1 |
| 2002 | 0 | 0.62 | 1 |
| 2003 | 0 | 0.65 | 1 |
| 2004 | 0 | 0.68 | 1 |
| 2005 | 0 | 0.69 | 1 |
| 2006 | 0 | 0.70 | 1 |
| 2007 | 0 | 0.70 | 1 |
| 2008 | 0 | 0.70 | 1 |
| 2009 | 0 | 0.71 | 1 |
| 2010 | 0 | 0.70 | 1 |
| 2011 | 0 | 0.70 | 1 |
| 2012 | 0 | 0.70 | 1 |
| 2013 | 0 | 0.71 | 1 |
| 2014 | 0 | 0.72 | 1 |
| 2015 | 0 | 0.73 | 1 |
| 2016 | 0 | 0.74 | 1 |
| 2017 | 0 | 0.75 | 1 |
| 2018 | 0 | 0.76 | 1 |

Table 7.8. Survey catch numbers-at-age and c.v.
Northern Irish Groundfish Q1

| year | c.v. | 1 | 2 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0.78 | 138.12 | 648.76 | 44.60 | 10.42 |
| 1994 | 0.34 | 1380.43 | 109.71 | 120.27 | 8.45 |
| 1995 | 0.68 | 700.73 | 386.15 | 20.03 | 10.78 |
| 1996 | 0.42 | 1106.13 | 329.28 | 111.67 | 1.39 |
| 1997 | 0.64 | 537.30 | 415.84 | 66.72 | 21.39 |
| 1998 | 0.84 | 169.36 | 769.23 | 56.87 | 11.98 |
| 1999 | 0.86 | 49.50 | 253.08 | 241.87 | 15.29 |
| 2000 | 0.65 | 629.60 | 101.053 | 34.58 | 33.01 |
| 2001 | 0.89 | 406.68 | 561.44 | 18.44 | 5.78 |
| 2002 | 0.64 | 662.16 | 253.31 | 333.54 | 0 |
| 2003 | 0.54 | 73.87 | 1079.20 | 104.05 | 32.70 |
| 2004 | 0.75 | 216.96 | 171.96 | 88.62 | 5.38 |
| 2005 | 0.76 | 63.53 | 225.07 | 29.41 | 27.96 |
| 2006 | 0.63 | 169.99 | 130.75 | 58.30 | 2.52 |
| 2007 | 0.95 | 164.35 | 124.39 | 30.60 | 5.15 |
| 2008 | 0.90 | 40.66 | 217.15 | 13.02 | 5.17 |
| 2009 | 0.76 | 144.00 | 59.00 | 33.00 | 9.00 |
| 2010 | 0.82 | 1022.12 | 208.96 | 14.66 | 2.26 |
| 2011 | 0.49 | 353.98 | 414.69 | 46.01 | 2.26 |
| 2012 | 0.81 | 161.90 | 222.82 | 99.27 | 14.25 |
| 2013 | 0.81 | 276.59 | 213.68 | 60.08 | 1.49 |
| 2014 | 0.63 | 314.41 | 222.80 | 53.29 | 13.66 |
| 2015 | 0.84 | 78.96 | 719.35 | 69.19 | 8.56 |
| 2016 | 1.06 | 349.20 | 175.00 | 148.30 | 10.70 |
| 2017 | 0.77 | 69.8 | 445.20 | 57.80 | 12.60 |
| 2018 | 1.26 | 138.1 | 50.50 | 62.60 | 0 |

## Northern Irish Groundfish Quarter 4

| year | c.v. | 0 | 1 | 2 |
| :---: | :---: | :---: | :---: | :---: |
| 1991 | 0.57 | 1109.37 | 50.06 | 47.60 |
| 1992 | 0.71 | 553.23 | 146.44 | 0.76 |
| 1993 | 0.45 | 1672.49 | 25.44 | 10.44 |
| 1994 | 0.38 | 1206.80 | 33.32 | 0 |
| 1995 | 0.60 | 486.65 | 50.15 | 6.54 |
| 1996 | 0.82 | 1322.20 | 97.19 | 0 |
| 1997 | 0.55 | 376.51 | 163.9 | 5.72 |
| 1998 | 0.75 | 58.47 | 32.48 | 9.49 |
| 1999 | 0.68 | 301.64 | 2.03 | 0 |
| 2000 | 0.72 | 506.79 | 109.91 | 0 |
| 2001 | 0.55 | 487.89 | 37.68 | 12.53 |
| 2002 | 0.86 | 161.45 | 29.4 | 0 |
| 2003 | 0.76 | 578.97 | 23.71 | 0 |
| 2004 | 0.82 | 706.13 | 107.72 | 17.28 |
| 2005 | 0.73 | 130.20 | 1.47 | 6.58 |
| 2006 | 1.22 | 86.99 | 0 | 2.98 |
| 2007 | 0.62 | 17.28 | 17.28 | 0 |
| 2008 | 1.09 | 213.62 | 6.1 | 0 |
| 2009 | 0.83 | 171.80 | 2.98 | 0 |
| 2010 | 0.82 | 92.48 | 53.86 | 3.05 |
| 2011 | 0.75 | 107.05 | 1.69 | 6.37 |
| 2012 | 0.72 | 321.82 | 32.79 | 20.33 |
| 2013 | 0.78 | 41.67 | 79.95 | 20.66 |
| 2014 | 0.78 | 0 | 55.35 | 39.15 |
| 2015 | 0.57 | 224.27 | 0 | 55.42 |
| 2016 | 0.83 | 14.98 | 0 | 181.79 |
| 2017 | 0.68 | 429.50 | 44.60 | 10.60 |
| 2018 | 1.42 | 68.50 | 112.60 | 0 |

## UK FSP

| year | c.v. | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0.4 | 0.43 | 1.41 | 0.99 | 0.08 | 0.03 |
| 2006 | 0.4 | 0.54 | 2.81 | 0.43 | 0.10 | 0.01 |
| 2007 | 0.4 | 0.61 | 1.32 | 0.59 | 0.06 | 0.06 |
| 2008 | 0.4 | 0.22 | 0.82 | 0.15 | 0.08 | 0.02 |
| 2009 | 0.4 | 0.17 | 1.15 | 0.38 | 0.10 | 0.02 |
| 2010 | 0.4 | 0.74 | 0.45 | 0.47 | 0.13 | 0.02 |
| 2011 | 0.4 | 0.41 | 1.68 | 0.14 | 0.10 | 0.04 |
| 2012 | 0.4 | 0.36 | 2.30 | 0.80 | 0.07 | 0.02 |
| 2013 | 0.4 | 0.84 | 1.88 | 1.35 | 0.37 | 0.06 |
| 2014 |  |  |  |  |  |  |
| 2015 | 0.4 | 0.60 | 2.04 | 1.17 | 0.26 | 0.05 |
| 2016 | 0.4 | 1.00 | 6.39 | 1.43 | 0.41 | 0.03 |
| 2017 | 0.4 | 3.06 | 2.85 | 3.84 | 1.01 | 0.23 |
| 2018 | 0.4 | 0.43 | 3.73 | 0.61 | 0.63 | 0.15 |

## MIKNET survey

| Year | c.v. | 0 |
| :---: | :---: | :---: |
| 1994 | 0.7 | 57.4 |
| 1995 | 0.7 | 6.9 |
| 1996 | 0.7 | 66.3 |
| 1997 | 0.7 | 5.7 |
| 1998 | 0.7 | 0.1 |
| 1999 | 0.7 | 26.2 |
| 2000 | 0.7 | 6.1 |
| 2001 | 0.7 | 9.6 |
| 2002 | 0.7 | 3.4 |
| 2003 | 0.7 | 3.2 |
| 2004 | 0.7 | 25.8 |
| 2005 | 0.7 | 11.4 |
| 2006 | 0.7 | 9 |
| 2007 | 0.7 | 0.0 |
| 2008 | 0.7 | 0.8 |
| 2009 | 0.7 | 23.6 |
| 2010 | 0.7 | 5.7 |
| 2011 | 0.7 | 1.4 |
| 2012 | 0.7 | 10.6 |
| 2013 | 0.7 | 42.6 |
| 2014 | 0.7 | 8.2 |
| 2015 | 0.7 | 80.4 |
| 2016 | 0.7 | 0.0 |
| 2017 | 0.7 | 10.6 |
| 2018 | 0.7 | 20.65 |

Table 7.9. Relative Biomass and Harv est rate used for the advice.

| Year | Relative Biomass index | Relative Harvest Rate |
| :---: | :---: | :---: |
| 1993 | 1.26 | 2.9 |
| 1994 | 1.06 | 2.4 |
| 1995 | 0.79 | 2.8 |
| 1996 | 1.09 | 2.2 |
| 1997 | 1.26 | 2.2 |
| 1998 | 1.48 | 1.69 |
| 1999 | 1.40 | 1.65 |
| 2000 | 0.80 | 1.14 |
| 2001 | 1.16 | 0.93 |
| 2002 | 2.2 | 0.57 |
| 2003 | 2.2 | 0.32 |
| 2004 | 0.76 | 0.81 |
| 2005 | 0.73 | 0.71 |
| 2006 | 0.48 | 0.99 |
| 2007 | 0.42 | 0.94 |
| 2008 | 0.41 | 0.81 |
| 2009 | 0.36 | 0.72 |
| 2010 | 0.70 | 0.56 |
| 2011 | 0.95 | 0.20 |
| 2012 | 0.78 | 0.52 |
| 2013 | 0.82 | 0.20 |
| 2014 | 0.84 | 0.22 |
| 2015 | 1.63 | 0.09 |
| 2016 | 1.05 | 0.06 |
| 2017 | 0.97 | 0.07 |
| 2018 | 0.35 | 0.34 |



Figure 7.1. Landings (grey) and discards- (white) at-age in total weight and numbers from 1992 to 2018.


Figure 7.2. Discard proportions-at-age 1995-2018.


Figure 7.3. Weight-at-age, ages 1-6.


Figure 7.4. Log ratio of ages in commercial catches.


Figure 7.5. Log-standardised age distribution in survey indices.


Figure 7.6. Survey age continuity.


Figure 7.7. Log ratio of cohorts in surveys.


Figure 7.8. Observed and predicted catches.



Figure 7.9. Commercial fleet catch-at-age residuals.


Figure 7.10. Index Fit.


Figure 7.11. Index fit at-age.


Figure 7.12. Index residuals catch-at-age (NIMIK is not included as only targets age 0 group.


Figure 7.13. Cod in ICES Division 7.a: ASAP Fishery selectivity-at-age, Block 1: 1968-1999, Block 2: 2000-today.


Figure 7.14. Cod in ICES Division 7.a: ASAP Index selectivity-at-age, Index1: NIGFSQ1, Index 2: NIGFSQ4, Index3: UK-FSP, Index 4: Miknet.


Figure 7.15. Model RMSE fit.


Figure 7.16. Estimated stock numbers-at-age.


Figure 7.17. ASAP retrospective summary 2011-2017.


Figure 7.18. Catches and landings in tons, harvest rate and biomass index as used for the advice. The two red lines indicate the 2 over 3 rule values.

## 6 Cod in Divisions 7.e-k (Eastern English Channel and southern Celtic Seas)

## Full analytical assessment

This stock has been benchmarked at WKROUND in February 2012. XSA was kept as the assessment model. Data, assessment and forecast procedure are detailed in the stock annex.

In 2019, the assessment was not of sufficient quality to be retained as a category 1 assessment. The basis for the advice is the ICES precautionary approach (category 3 assessment).

Latest ICES advices in 2018 and 2019
2018: "ICES advises that when the MSY approach is applied, there should be zero catch in 2019."
2019: "For Cod in divisions $7 . e-k$, ICES advises that when the precautionary approach is applied, there should be zero catch in 2020."

### 6.1 General

## Stock description and management units

The TAC is set for ICES areas $7 . \mathrm{b}-\mathrm{c}, 7 . \mathrm{e}-\mathrm{k}, 8,8,10$, and CECAF 34.1.1(1), excluding 7.d. This is representative of the stock area as the cod population in 7. d is more relevant to the North Sea population. However, landings from 7.bc are not included in the assessment area.

## Management applicable in 2018 and 2019

## TAC 2018 (Council regulation 2018/120)

| Species: | Cod Gadus morhua |  | Zone: | $7 \mathrm{~b}, 7 \mathrm{c}, 7 \mathrm{e}-\mathrm{k}, 8,9$ and 10 ; Union waters of CECAF 34.1.1 <br> (COD/7XAD34) |
| :---: | :---: | :---: | :---: | :---: |
| Belgium |  | 121 |  |  |
| France |  | 1984 |  |  |
| Ireland |  | 757 |  |  |
| The Netherlands |  | 0 |  |  |
| United Kingdom |  | 214 |  |  |
| Union |  | 3076 |  |  |
| TAC |  | 3076 |  | Analytical TAC <br> Article 12(1) of this Regulation applies |

TAC 2019 (Council regulation 2019/124)


Since 2005, ICES rectangles 30E4,31E4, and 32E3 have been closed during the first quarter (Council Regulations 27/2005, 51/2006, and 41/2007, 40/2008, and 43/2009).

Technical measures applied to this stock are a minimum mesh size (MMS) for beam and otter trawlers in Subarea 7 and a minimum landing size (MLS) of 35 cm .

## Fishery

Landings data used by the WG are summarised in Table 8.1 and the Figure 8.1 provides historical landings by countries. In 2018, the landings are 1385 t and represent an uptake of $45 \%$ of the agreed TAC.

TAC uptake varies among countries. Belgium uses $40 \%$ of their TAC. Ireland takes $93 \%$ of its TAC. United Kingdom uses $61 \%$ of its TAC and France undershooting its TAC to only $25 \%$. United Kingdom catches mainly comes from UK England. The low uptake rate for France is the consequence of the mixed nature of its fisheries. Cod is no longer a target species but as bycatch in haddock and whiting dedicated fisheries.

Given the rapid growth of cod in this area, discards are mostly composed of one and two year old fish. Since 2011, quotas were not restricted and the discard rate has been stable around 10$15 \%$. Discards estimate for 2018 is $180 t$, which corresponds to a discards rate of $11.5 \%$. Discards rate by countries are $13.5 \%, 12.3 \%, 9.7 \%$ and $13 \%$ for Belgium, Ireland, France and United Kingdom respectively.

Captures (in tonnes) and TAC uptake percentage by catch category and country.

| Country | CatchCategory | Tons | TAC | TAC_Uptake |
| :--- | :--- | :--- | :--- | :--- |
| Ireland | BMS landing | 0 | NA | NA |
| Belgium | Discards | Discards | 5.6 | NA |

Total catches (i.e. landings and discards), discards and discards rate by country. Total catches and discards are in tonnes.

| Country | Catches (L+D) | Discards | Discard_rate |
| :--- | :---: | :---: | :---: |
| Belgium | 56.2 | 7.6 | 13.5 |
| France | 553.1 | 53.8 | 9.7 |
| Ireland | 805 | 99.1 | 12.3 |
| Netherlands | 0.7 | 0.2 | 28.6 |
| United Kingdom | 149.6 | 19.5 | 13 |

Cod is mainly caught in area 27.7.g, followed by areas 27.7.h, 27.7.e and 27.7.j respectively. No landings are reported in 27.7.k and few in 27.7.j2 (Figure 8.2).

France is fishing in all areas but most of its landings are taking in 27.7.g. Ireland and Belgium are fishing in 27.7.g and UK in 27.7.e. For each country, landings distribution in the Celtic Sea is similar to previous years.
In Celtic Sea, cod is mainly caught by OTB_DEF_100-119_0_0_all, OTT_DEF_100-119_0_0_all and OTB_DEF_70-99_0_0_all métiers. Otter trawls landings represent $76 \%$ of the total landings. Beamers (i.e. TBB_DEF_70-99_0_0_all) also contribute to cod landings but in a lower proportion ( $10 \%$ of the total landings).

Discards rate in weight varies among métiers depending on gear, mesh size range and targeted species. Most of the discards are realised by otter trawlers and beamers, $60 \%$ and $22 \%$ respectively.

The group advises to follow métier definition specified in the Appendix 2 of the ICES data call to reduce the number of métier upload in InterCatch. Métier which contribute to less than $1 \%$ of the landings are included in the MIS_MIS_0_0_0_HC métier.

Captures amounts (tonnes) and percentages by catch category and métier.

| Gears | Landings | Discards |
| :--- | :---: | :---: |
| Otter_trawls | 1043 | 102 |
| Beamers | 137 | 37 |
| Seiners | 96 | 31 |
| Gillnets | 81 | 0 |
| Others | 10 | 0 |

Information from the Industry
No specific information was reported to the group in 2019.

### 6.2 Data

## InterCatch procedure

Since 2013, international landings and discards data are uploaded in InterCatch. Discards are raised for unreported strata to estimate total discards in weight.

Unsampled strata of landings and discards (number-at-age) are filled in using an allocation procedure. Information on national and international assumptions madeby data providers and submitters at the national level and allocation grouping used in IC are available on SharePoint To ensure the consistency of data processing at international level, the same rules are applied each year for the allocation procedure: fill unsampled strata using as much as possible the same métier and quarter, regardless of area and country. Unsampled BMS landings and Logbook Registered Discards are filled in using discards data employing as much as possible the same métier and quarter, regardless of area and country.
One of the ToRs proposed for thenext benchmark is tostreamlining data compilation procedures for fishery-dependent data of the three main gadoids species (cod, haddock and whiting). General raising protocol would then be added to the stock annex.

## Landings

Length distributions of 2018 landings provided by countries for sampled strata and quarter are shown in Figures 8.3 a-d.

Age distribution of 2018 catches (i.e. landings and discards) is illustrated in the Figure 8.4. It is noticeable that this stock has always been composed of few age classes, even though Celtic Sea cod can live up to ten years. While the catch was mainly composed of age 2 over the period 20052008, the strong 2009 year class has contributed strongly to the catch at older ages in recent years: $63 \%$ in number in 2012 at-age $3,36 \%$ at age 4 in 2013 (Table 8.2.a and 8.2.b). In 2014, high recruitment has been observed resulting in an increasing proportion of age 1 fish in the landings (53\%), age 2 accounts for $22 \%$ of the landings. In 2015, landings are dominated by fish of age 2, in 2016 landings are dominated by fish of age 3 and in 2017 landings was mostly composed by cod of
age 2. In $2018,20 \%$ of the landings was fish of age 1, $35 \%$ of fish of age 2 and $31 \%$ of fish of age 3.

## Discards

The landings/discards pattern is known to be strongly variable between fleets and years due to métier, recruitment intensity, TACs constraints and mixed fisheries concerns.

In 2009, age 1 individuals ( $30-45 \mathrm{~cm}$, Mahé et al. 2016) were mainly discarded. In 2010, most of them were landed. In 2011, ages 1 and 2 represents respectively $51 \%$ and $46 \%$ of the total discards in numbers for all fleets. Due to the low TAC relative to the high magnitude of recruitment in 2009 and 2010, all countries had unusually high discard rates in 2011, generally $70 \%$ by weight was made up of fish above the minimum landing size (MLS, i.e. 35 cm for Celtic Sea cod). The high-graded fish from the French fishery have been added to the landings in 2003-2011. In 2014, total amount of discards was 740 t ( 639 t imported +101 t raised), giving a discard rate of $19 \%$. This discards rate was higher than the average $10 \%$ and mostly consisted of undersized fish from the strong 2013 year class (fish of age 1 in 2014). In 2015, the total amount of discard was 565 t ( 250 t sampled and uploaded in InterCatch and 309 t resulting from the raising procedures), giving a discard rate by weight of $12 \%$, which is considered the usual discard rate for this species in the mixed fisheries. High-grading in 2015 (discards of fish above Minimum conservation size) was low. In 2016, the total amount of discards was 220 t ( 154 t sampled and uploaded in InterCatch and $52 t$ resulting from the raising procedures), giving a discard rate by weight of $6.3 \%$. In 2017, the total amount of discards was 117 t ( 47 t sampled and uploaded in InterCatch and 62 t resulting from the raising procedures), giving a discard rate by weight of $5 \%$, which is considered lower than average. They are mainly composed of age 1 fish (Figure 8.4).

Length distributions of 2017 discards provided by countries for sampled strata and quarter are shown Figure 8.3a-d. In recent years, due to quota constraints at vessels levels, length distribution of discards for the UK fleet have shown high-grading pattern (cod being a non-target species). However, this fleet has little contribution toboth, landings and discards quantities and this was no more reported in 2017.

In 2018, discards are mostly composed of fish of 1 year (Figure 8.4).

## Biological

Landings numbers-at-age (before and after the SOP correction), catch (i.e. landings) and stock weights-at-age are given respectively in Tables 8.2.a, 8.2.b, 8.3 and 8.4.

Biological parameters are described in the stock annex and are unchanged since the 2012 WKROUND benchmark. Celtic Sea cod are very fast growing and early maturing compared with more northern cod stocks.

## Commercial Ipue

Tables 8.5 a-c gather the values of landings, fishing effort and lpue data series for the French (a), Irish (b) and UK fleets (c). Figures 8.5 a-c illustrate the trends of lpue and effort by country.

A general decrease in the lpue trend is observed in almost all series between 1990 and 2004, where the TAC began to be constraining. From that point, the lpue seemed to stabilize, or even to increase if high-grading is taken into account. The strong 2009 year class resulted in an increase of lpue for all fleets between 2010 and 2012. Different features are observed in the effort timeseries. The métiers showing the highest levels of cod directed effort have decreased significantly in the last 5-10 years until 2010. Since then, effort has gone up again until 2013 following the increased of TAC possibilities.

Since 2013, French fishing effort and lpue have decreased (Figure 8.5a). Effort of Irish fleet targeting gadoids (i.e. Otter trawl27.7.g) remains at a high level as a consequence of mixed fisheries interaction with increased whiting and haddock fisheries opportunities (Figure 8.5b). In the meantime the spawning-stockbiomass(SSB) is low, as suchlpue is decreasing since 2013(Figure 8.5b). In 2018, Otter trawl Irish 27.7.g lpue has increased. Effort of the UK trawl fleet in 27.7.e-k shows a decreasing trend (down to zero in 2016) and increases since then, while beam trawl effort in 27.7.e-k relatively stable in recent years (Figure 8.5c).
Remark: The UK English and Welsh effort data are only reliable for vessels over 12 metres registered length, and therefore has always been provided to working groups for vessels greater than 12 metres. The fleet of over 12 meter vessels has been declining gradually over the years, until in 2016 no effort recorded from this fleet. The zero figures provided for 2016 have been checked and are correct (Figure 8.5c).

## Surveys and commercial tuning fleet

Table 8.6 presents the survey data series. Two ongoing surveys, both part of the DCF, IBTS Q4 (FR-EVHOE \& IR-GFS7gj combined) are used to assess this stock (see details in the stock annex and modification based on 2014 WKCELT benchmark).
The historical time-series of age structure of the commercial tuning index (OTDEF French fleet for quarter 2,3 and 4) and the survey index are shown in Table 8.6 and Figure 8.6.
In 2017, the French EVHOE survey was not conducted due to technical difficulties at the beginning of the survey. The IR-FR combined tuning index used in the assessment is only composed of Irish data for 2017. The Irish survey covered additional stations normally undertaken by the EVHOE survey.

## Dataissues

No important issues were reported this year.
Remark: When for a métier/strata landings are upload annually, there are is noinformation available in InterCatch to split the annual landings into quarterly landings and therefore the associated age composition and mean weight-at-age. As a result, when extracting quarter 1 versus quarter 2,3 and, 4 data to inform on mean weight of the stock and the catch for the assessment, these data are not used.

### 6.3 Stock assessment

Model used: XSA.

## Final update assessment (XSA)

The final assessment was run with the same settings as established by WKROUND 2012 and described in the stock annex. Discards are not included in the assessment.

VPA. 95 software was run in parallel to the FLRXSA R script to fully validate the assessment.
XSA residuals and diagnostics donot highlight any problem regarding the input data and model fit (Table 8.7 and Figure 8.7). Outputs from the assessment are reported in Tables 8.8-10 and in Figures 8. 7-11.

The comparison of runs with and without tuning indices indicates that the majority of the information comes from the catch-at-age matrix (Figure 8.11b). The information contains in both indices seems contradictory. In recent years, survey indices contains little information due to low recruitment levels.

In 2019, the assessment shows a substantial downward revision in SSB and recruitment in recent years and substantial upward revision in F. Comparing this year's assessment results with last year and looking at the retrospective analysis, a rescaling in spawning-stock biomass (SSB), recruitment and $F$ is evident (Figure 8.11a).

Mohn's rho analysis (i.e. a measure of the relative difference between an estimate from an assessment with a truncated time-series and an estimate of the same quantity from an assessment using the full time-series) resulted in values of -0.38 for $F_{\text {bar }}(2-5), 0.48$ for SSB and 0.27 for recruitment. $F$ is revised up every year of since 2012. Compared to last year results, $F$ of age 2 to 5 was resided up by a factor of 1.6 to 2.08 .

ICES considers a value greater than 0.20 to be unacceptably high. However, the stock is maintained in category 1.
The variability of cod recruitment over years is partially responsible the retrospective patterns, however, substantial changes in assessment between years may be due to strong year classes dominating the fishery in recent years, the unexpected disappearance of fish of older age, and discards being only partially included in the assessment.

## State of the stock

Table 8.8 and 8.9 summarise the estimated fishing mortality-at-age and the stock numbers-atage, respectively. The stock summary is reported in Table 8.10 and Figure 8.10.
Catches are around 5000 t since 2000 (Figure 8.10), with some higher catches following strong recruitments. Reliable discard estimates are only available since 2011 and range between 150 and 1000 t depending on the interplay between recruitment dynamics and TAC constraints.
Recruitment has been highly variable over time with occasional very high recruitment followed by period of low recruitments. Since 2012, recruitment has been very weak with the exception of the 2014 year class, which is above average (Table 8.10 and Figure 8.10). Recruitment estimated in 2017 and 2018 are remarkably low (Table 8.10 and Figure 8.10).

Spawning-stock biomass (SSB) is well below MSY $B_{\text {trigger }}$ and $\mathrm{B}_{\mathrm{pa}}$ (both at 10300 t ) since 2000, with the exception of 2012 as the consequence of a very good recruitment year (Table 8.10 and Figure 8.10, ICES, 2012). Since 2004, SSB is also below Blim (7300 t), except during the 2011-2013 period (Table 8.10 and Figure 8.10, ICES, 2012).

Fishing mortality has been above $\mathrm{F}_{\text {msy }}$ for the entire time-series but has been decreasing between 2000 and 2010 and increased again after (Table 8.10 and Figure 8.10. Fishing mortality fluctuated around $\mathrm{Flim}_{\text {lim }}$ in recent years. In 2018, F was estimated at 0.83 that is above $\mathrm{F}_{\lim }(0.80), \mathrm{F}_{\mathrm{pa}}(0.58)$ and well above FMSY (0.35) (ICES, 2016a; 2016b).

### 6.4 Short-term projections

Assumptions made for the short-term projections are described in the table below.

Assumptions made for the interim year and in the forecast.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| F ages 2-5 (2018) | 0.865 | $\mathrm{~F}_{\text {sq }}=\mathrm{F}_{\text {average }}(2016-2018)$ |
| SSB (2020) | 1664 tonnes | Fishing at $\mathrm{F}=0.865$ |
| $\mathrm{R}_{\text {age 1 }}(2019-2020)$ | 2166 thousand | 25th quantile of the recruitment time-series (1971-2018) |
| Catch (2019) | 1321 tonnes | Landings + estimated discards |
| Wanted catch (2019) | 1228 tonnes | Fishing at F=0.865 |
| Unwanted catch (2019) | 93 tonnes | Average discards rate (2016-2018) $=7.58 \%$ |

F-at-age (range 2 to 5 ) is based on the three past year average. Recruitment (age 1 ) is assumed as 25 th quantile of the recruitment time-series (1971-2018). This option was preferred to the geometric mean in recent years as the previous year's SSB are low. Discards rate is based on the average rate of the past three years. Weights-at-age is estimated on the basis of three year average. No TAC constraint was applied.

The inputs to the short-term predictions and their outputs are presented in Table 8.11 and 8.12, respectively.

In 2019, under the forecast assumption, wanted catch (i.e. the estimated landed fish) are predicted to be 1228 t (which is less than the TAC and but higher than the 2018 ICES zero catch advice) (Table 8.13) and unwanted catch are predicted to be 93 t .

SSB is predicted to be 1664 t in 2020 which would still be above $\mathrm{Blim}_{\lim }(7300 \mathrm{t})$, MSY Brrigger and $\mathrm{B}_{\mathrm{pa}}$ (10 300 t) (Table 8.13).

The forecasts are sensitive to the recruitment assumption that contributes to $58 \%$ of the landings in 2020 and 52\% of the projected SSB in 2020 (Table 8.12 and Figure 8.13).

### 6.5 Medium-term projection

No medium-term projections were carried out.

### 6.6 Biological reference points

The reference points has been estimated using the agreed ICES guidelines (ICES, 2016).

Reference points

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY <br> approach | MSY $\mathrm{B}_{\text {trigger }}$ | 10300 tonnes | $\mathrm{B}_{\mathrm{pa}}$ | ICES (2012) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.35 | Segmented regression with $\mathrm{B}_{\text {lim }}$ (EqSim) | ICES (2016) |
|  | $\mathrm{F}_{\text {MSY lower }}$ | 0.23 | F at 95\% of MSY below $\mathrm{F}_{\text {MSY }}$ | ICES (2016) |
|  | $\mathrm{F}_{\text {MSY upper }}$ | 0.55 | F at 95\% of MSY above $\mathrm{F}_{\text {MSY }}$ | ICES (2016) |
| Precautionary approach | $\mathrm{B}_{\text {lim }}$ | 7300 tonnes | $\mathrm{B}_{\text {loss, }}$, lowest observed SSB (1976), rounded value | ICES (2012) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 10300 tonnes | $\mathrm{B}_{\lim } \times 1.4$ | ICES (2012) |
|  | $\mathrm{F}_{\text {lim }}$ | 0.80 | Segmented regression with $\mathrm{Bl}_{\text {lim }}$ (EqSim) | ICES (2016) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.58 | $\mathrm{F}_{\text {lim }} / 1.4$ | ICES (2016) |
| Management plan | $\mathrm{SSB}_{\text {MGT }}$ | Not applicable |  |  |
|  | $\mathrm{F}_{\text {MGT }}$ | Not applicable |  |  |

### 6.7 Management plans

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to demersal stocks including cod in ICES divisions 7.e-k.

## Uncertainties and bias in assessment and forecast

Cod, haddock and whiting in the Celtic Sea will be benchmarked together in 2020. The focus of the benchmark would be on streamlining data compilation procedures for fishery dependent and survey data. This will improve transparency and diagnostics surrounding commercial tuning fleets and surveys. The benchmark should also review the assessment methods and diagnostics given the potential for changes in selectivity in the commercial fishery. The benchmark should also investigate mixed fisheries and multi-species interactions as well as environmental drivers that may be impacting on growth and recruitment of all three species.

Issues that might causes retrospectives bias are:
i. The non-inclusion of undersized discards (and high-grading in recent years) in the assessment. However, high-grading is estimated at a very low level in recent year because the TACs were not constraining (undershoot TACs).
ii. Sensitivity analysis of the assessment to commercial tuning series calculation should be investigating during the next benchmark process.

### 6.8 Management considerations

Management scenario are summarised in the Table 8.14.

No management scenario can bringSBB above $B_{p a}$ in 2021 under the current recruitment assumption. The strong retrospective pattern implies that the current $F$ estimates might be uncertain. Forecasts are sensitive to the assumption on recruitment as the landings are usually composed of a high proportion of age 2 fish (and age 1 fort discards).

The recent technical measures introduced in the Celtic Sea (square mesh panels) are not expected to significantly reduce catches of Celtic Sea cod or improved the selection pattern. This is because of the fast growth rate of Celtic Sea cod (age 2 fish range between 40 and 70 cm , Mahé et al., 2016).

The strong upward revision in F in previous year's assessment implies that the stock has never been fished at FMSY which could explain why SSB is still below MSY Btrigger. Additionally, mixed fisheries issues could be responsible for maintaining F at high level, as other gadoids fishing opportunities are higher. In this context, cod is no longer a target species but can be considered as bycatch in the fleet targeting haddock, whiting and Nephrophs.

Historical information on management consideration can be found in the stock annex.
Given that SSB is estimated to be well below Blim and is likely to remain so in 2019 and 2020, the ICES framework for category 3 stocks was not applied (ICES, 2012). The advice is for zero catch for 2020. Recent recruitment has been low and the short-term outlook is very dependent recent recruitment for this stock.

### 6.9 References

ICES. 2012. Report of the Working Group on the Celtic Seas Ecoregion (WGCSE), 9-18 May 2012, Copenhagen, Denmark. ICES CM 2012/ACOM:12.

ICES. 2016. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

Mahé K., Dufour J.L., Brown D., Smith J., Beattie, S. and Woods F. Working paper : Cod (Gadus morhua) in the Celtic Sea otolith exchange 2016.

Table 8.1. Cod in divisions 7.e-k. History of official commercial landings presented by country and used by the Working Group. All weights are in tonnes.

| Year | Belgium | France | Ireland | UK | Others | Total | High-graded dis- <br> card estimates |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Discard es- <br> timates | Landings taken or <br> reported in 33E2 <br> and 33E3 *** |  |  |  |  |  |  |
| 1971 | NA | NA | NA | NA | NA | 5782 | NA |
| 1972 | NA | NA | NA | NA | NA | 4737 | NA |
| 1973 | NA | NA | NA | NA | NA | 4015 | NA |


| Year | Belgium | France | Ireland | UK Others | Total | High-graded dis- <br> card estimates | Discard es- <br> timates | Landings taken or <br> reported in 33E2 <br> and 33E3 *** |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 528 | 8096 | 1718 | 1089 | 0 | 11431 |  | NA |  |  |
| 1999 | 326 | 5488 | 1883 | 897 | 0 | 8594 |  | NA |  | NA |

French high-grading estimates from self-sampling program. ${ }^{* *}$ International high-grading estimates.
*** Included in Ireland landings estimates. Landings in the south of Division 7.a (33E2 and 33E3) are included in the assessment and are considered to be part of the stock.

Table 8.2a. Cod in divisions 7e-k. Landings number-at-age (in thousands) (note: 2011 values represent actual catch) InterCatch outputs before SOP correction.

| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 | Age 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 725 | 461 | 557 | 96 | 35 | 17 | 5 | 5 | 1 | 0 |
| 1972 | 4 | 774 | 110 | 205 | 45 | 26 | 11 | 5 | 1 | 0 |
| 1973 | 332 | 239 | 346 | 60 | 74 | 17 | 6 | 4 | 1 | 0 |
| 1974 | 1 | 224 | 40 | 118 | 38 | 37 | 18 | 4 | 14 | 0 |
| 1975 | 673 | 136 | 185 | 61 | 105 | 20 | 20 | 12 | 1 | 0 |
| 1976 | 51 | 1456 | 61 | 107 | 11 | 22 | 2 | 4 | 1 | 0 |
| 1977 | 25 | 416 | 236 | 15 | 60 | 2 | 2 | 5 | 10 | 0 |
| 1978 | 197 | 497 | 129 | 116 | 20 | 34 | 6 | 8 | 4 | 2 |
| 1979 | 438 | 357 | 263 | 68 | 104 | 19 | 24 | 5 | 2 | 1 |
| 1980 | 609 | 1213 | 285 | 175 | 52 | 55 | 14 | 0 | 0 | 0 |
| 1981 | 315 | 3086 | 811 | 153 | 41 | 20 | 10 | 2 | 0 | 0 |
| 1982 | 76 | 1157 | 888 | 169 | 36 | 19 | 4 | 1 | 0 | 0 |
| 1983 | 1285 | 529 | 540 | 424 | 77 | 21 | 5 | 5 | 1 | 0 |
| 1984 | 737 | 1210 | 134 | 97 | 94 | 22 | 3 | 2 | 0 | 0 |
| 1985 | 726 | 1245 | 465 | 61 | 40 | 47 | 12 | 2 | 1 | 0 |
| 1986 | 651 | 1303 | 673 | 254 | 30 | 31 | 17 | 0 | 0 | 0 |
| 1987 | 2741 | 946 | 448 | 250 | 62 | 20 | 11 | 4 | 0 | 0 |
| 1988 | 1830 | 5443 | 320 | 133 | 46 | 21 | 4 | 2 | 2 | 0 |
| 1989 | 666 | 2639 | 2483 | 149 | 77 | 18 | 8 | 2 | 1 | 0 |
| 1990 | 360 | 846 | 1006 | 663 | 79 | 21 | 8 | 6 | 2 | 0 |
| 1991 | 1377 | 1034 | 229 | 330 | 203 | 48 | 11 | 3 | 0 | 0 |
| 1992 | 1434 | 2601 | 329 | 64 | 70 | 53 | 16 | 1 | 0 | 0 |
| 1993 | 274 | 2371 | 928 | 79 | 24 | 19 | 14 | 2 | 0 | 0 |
| 1994 | 1340 | 692 | 1199 | 258 | 27 | 10 | 11 | 6 | 0 | 0 |
| 1995 | 823 | 3320 | 310 | 284 | 73 | 13 | 2 | 3 | 0 | 0 |
| 1996 | 617 | 2248 | 1199 | 134 | 95 | 43 | 3 | 1 | 0 | 0 |
| 1997 | 1184 | 1870 | 951 | 297 | 48 | 22 | 6 | 0 | 0 | 0 |
| 1998 | 639 | 2545 | 641 | 254 | 99 | 36 | 6 | 2 | 0 | 0 |


| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 | Age 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 496 | 1141 | 756 | 158 | 59 | 36 | 9 | 5 | 0 | 0 |
| 2000 | 1693 | 464 | 419 | 169 | 44 | 17 | 12 | 2 | 0 | 0 |
| 2001 | 1091 | 2373 | 136 | 98 | 70 | 19 | 12 | 6 | 1 | 0 |
| 2002 | 210 | 2069 | 883 | 64 | 33 | 12 | 6 | 4 | 1 | 0 |
| 2003 | 103 | 556 | 827 | 217 | 15 | 9 | 6 | 1 | 0 | 0 |
| 2004 | 341 | 298 | 175 | 168 | 59 | 8 | 4 | 3 | 0 | 0 |
| 2005 | 295 | 664 | 138 | 52 | 45 | 11 | 2 | 0 | 0 | 0 |
| 2006 | 368 | 994 | 249 | 25 | 14 | 13 | 4 | 1 | 0 | 0 |
| 2007 | 491 | 1245 | 409 | 60 | 9 | 4 | 3 | 1 | 0 | 0 |
| 2008 | 123 | 769 | 312 | 101 | 24 | 4 | 3 | 1 | 0 | 0 |
| 2009 | 161 | 281 | 324 | 96 | 37 | 10 | 2 | 0 | 0 | 0 |
| 2010 | 532 | 434 | 122 | 91 | 42 | 9 | 2 | 0 | 0 | 0 |
| 2011 | 1516 | 3158 | 232 | 52 | 32 | 9 | 2 | 0 | 0 | 0 |
| 2012 | 35 | 489 | 1346 | 219 | 26 | 14 | 4 | 0 | 3 | 0 |
| 2013 | 110 | 195 | 433 | 451 | 65 | 21 | 6 | 0 | 0 | 0 |
| 2014 | 762 | 327 | 82 | 113 | 134 | 9 | 1 | 0 | 0 | 0 |
| 2015 | 37 | 1576 | 119 | 21 | 34 | 27 | 8 | 1 | 0 | 0 |
| 2016 | 137 | 89 | 579 | 33 | 6 | 10 | 17 | 1 | 0 | 0 |
| 2017 | 19 | 431 | 83 | 119 | 16 | 4 | 5 | 2 | 0 | 0 |
| 2018 | 88 | 152 | 135 | 12 | 34 | 3 | 1 | 1 | 0 | 0 |

Table 8.2b. Cod in divisions 7e-k. Landings number-at-age (in thousands) used in the assessment (note: 2011 values represent actual catch) - InterCatch outputs after SOP correction.

| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 725 | 461 | 557 | 96 | 35 | 17 | 11 |
| 1972 | 4 | 772 | 110 | 204 | 45 | 26 | 17 |
| 1973 | 331 | 239 | 345 | 60 | 74 | 17 | 11 |
| 1974 | 1 | 223 | 40 | 118 | 38 | 37 | 36 |
| 1975 | 674 | 136 | 185 | 61 | 105 | 20 | 33 |
| 1976 | 51 | 1460 | 61 | 107 | 11 | 22 | 7 |
| 1977 | 25 | 416 | 236 | 15 | 60 | 2 | 17 |
| 1978 | 196 | 496 | 129 | 116 | 20 | 34 | 20 |
| 1979 | 438 | 357 | 263 | 68 | 104 | 19 | 32 |
| 1980 | 609 | 1213 | 285 | 175 | 52 | 55 | 14 |
| 1981 | 315 | 3087 | 811 | 153 | 41 | 20 | 12 |
| 1982 | 77 | 1174 | 901 | 171 | 37 | 19 | 5 |
| 1983 | 1286 | 529 | 540 | 424 | 77 | 21 | 11 |
| 1984 | 736 | 1208 | 134 | 97 | 94 | 22 | 5 |
| 1985 | 733 | 1256 | 469 | 62 | 40 | 47 | 15 |
| 1986 | 651 | 1303 | 673 | 254 | 30 | 31 | 17 |
| 1987 | 2698 | 931 | 441 | 246 | 61 | 20 | 15 |
| 1988 | 1829 | 5441 | 320 | 133 | 46 | 21 | 8 |
| 1989 | 666 | 2640 | 2484 | 149 | 77 | 18 | 11 |
| 1990 | 356 | 838 | 996 | 656 | 78 | 21 | 16 |
| 1991 | 1377 | 1034 | 229 | 330 | 203 | 48 | 14 |
| 1992 | 1434 | 2601 | 329 | 64 | 70 | 53 | 17 |
| 1993 | 274 | 2373 | 929 | 79 | 24 | 19 | 16 |
| 1994 | 1340 | 692 | 1199 | 258 | 27 | 10 | 17 |
| 1995 | 823 | 3320 | 310 | 284 | 73 | 13 | 5 |
| 1996 | 617 | 2248 | 1199 | 134 | 95 | 43 | 4 |
| 1997 | 1185 | 1871 | 952 | 297 | 48 | 22 | 6 |
| 1998 | 640 | 2548 | 642 | 254 | 99 | 36 | 8 |


| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 497 | 1143 | 757 | 158 | 59 | 36 | 14 |
| 2000 | 1692 | 464 | 419 | 169 | 44 | 17 | 14 |
| 2001 | 1090 | 2371 | 136 | 98 | 70 | 19 | 19 |
| 2002 | 210 | 2068 | 883 | 64 | 33 | 12 | 11 |
| 2003 | 103 | 556 | 826 | 217 | 15 | 9 | 7 |
| 2004 | 341 | 298 | 175 | 168 | 59 | 8 | 7 |
| 2005 | 296 | 665 | 138 | 52 | 45 | 11 | 2 |
| 2006 | 368 | 995 | 249 | 25 | 14 | 13 | 5 |
| 2007 | 492 | 1246 | 409 | 60 | 9 | 4 | 4 |
| 2008 | 123 | 771 | 313 | 101 | 24 | 4 | 4 |
| 2009 | 161 | 281 | 324 | 96 | 37 | 10 | 2 |
| 2010 | 534 | 435 | 122 | 91 | 42 | 9 | 2 |
| 2011 | 1515 | 3156 | 232 | 52 | 32 | 9 | 2 |
| 2012 | 35 | 490 | 1349 | 219 | 26 | 14 | 7 |
| 2013 | 110 | 195 | 434 | 452 | 65 | 21 | 6 |
| 2014 | 747 | 320 | 80 | 111 | 131 | 9 | 1 |
| 2015 | 36 | 1518 | 115 | 20 | 33 | 26 | 9 |
| 2016 | 132 | 86 | 558 | 32 | 6 | 10 | 17 |
| 2017 | 18 | 406 | 78 | 112 | 15 | 4 | 7 |
| 2018 | 85 | 146 | 130 | 12 | 33 | 3 | 2 |

Table 8.3. Cod in divisions 7e-k. Catch (landings) weight-at-age.

| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 | Age 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0.908 | 2.193 | 4.831 | 7.464 | 9.669 | 11.784 | 13.862 | 15.494 | 16.195 | 16.315 |
| 1972 | 0.908 | 2.193 | 4.831 | 7.464 | 9.669 | 11.784 | 13.862 | 15.494 | 16.195 | 16.315 |
| 1973 | 0.908 | 2.193 | 4.831 | 7.464 | 9.669 | 11.784 | 13.862 | 15.494 | 16.195 | 16.315 |
| 1974 | 0.908 | 2.193 | 4.831 | 7.464 | 9.669 | 11.784 | 13.862 | 15.494 | 16.195 | 16.315 |
| 1975 | 0.908 | 2.193 | 4.831 | 7.464 | 9.669 | 11.784 | 13.862 | 15.494 | 16.195 | 16.315 |
| 1976 | 0.908 | 2.193 | 4.831 | 7.464 | 9.669 | 11.784 | 13.862 | 15.494 | 16.195 | 16.315 |
| 1977 | 0.908 | 2.193 | 4.831 | 7.464 | 9.669 | 11.784 | 13.862 | 15.494 | 16.195 | 16.315 |
| 1978 | 0.908 | 2.193 | 4.831 | 7.464 | 9.669 | 11.784 | 13.862 | 15.494 | 16.195 | 16.315 |
| 1979 | 0.908 | 2.193 | 4.831 | 7.464 | 9.669 | 11.784 | 13.862 | 15.494 | 16.195 | 16.315 |
| 1980 | 0.908 | 2.193 | 4.831 | 7.464 | 9.669 | 11.784 | 13.862 | 15.494 | 16.195 | 16.315 |
| 1981 | 0.945 | 1.549 | 4.385 | 7.565 | 9.060 | 12.750 | 13.822 | 19.232 | 19.232 | 19.232 |
| 1982 | 0.945 | 2.242 | 4.474 | 7.797 | 10.250 | 12.465 | 15.074 | 16.908 | 18.538 | 20.949 |
| 1983 | 0.979 | 2.525 | 4.961 | 7.457 | 9.965 | 12.010 | 14.767 | 17.643 | 19.131 | 19.131 |
| 1984 | 0.981 | 2.645 | 5.284 | 7.828 | 9.758 | 11.672 | 14.548 | 16.527 | 16.527 | 16.527 |
| 1985 | 1.001 | 2.637 | 5.521 | 8.082 | 10.407 | 11.469 | 13.448 | 16.658 | 20.853 | 20.853 |
| 1986 | 1.054 | 2.554 | 5.398 | 7.440 | 10.782 | 12.396 | 13.558 | 13.558 | 13.558 | 13.558 |
| 1987 | 0.909 | 2.504 | 5.264 | 8.089 | 10.447 | 13.574 | 15.029 | 16.229 | 16.229 | 16.229 |
| 1988 | 0.906 | 2.187 | 5.318 | 7.997 | 10.649 | 12.486 | 13.805 | 14.285 | 16.592 | 16.592 |
| 1989 | 0.844 | 2.013 | 4.706 | 7.638 | 9.438 | 12.917 | 12.479 | 15.407 | 16.683 | 16.683 |
| 1990 | 0.880 | 2.300 | 4.624 | 7.188 | 9.045 | 11.713 | 13.769 | 16.786 | 13.081 | 13.081 |
| 1991 | 0.905 | 2.135 | 4.987 | 6.738 | 8.865 | 10.809 | 13.768 | 15.478 | 15.478 | 15.478 |
| 1992 | 0.815 | 1.916 | 4.916 | 7.359 | 9.744 | 11.498 | 12.474 | 15.117 | 15.117 | 15.117 |
| 1993 | 0.871 | 2.043 | 4.508 | 6.866 | 8.431 | 10.942 | 12.147 | 13.646 | 16.530 | 16.530 |
| 1994 | 0.874 | 2.000 | 4.492 | 7.926 | 10.092 | 12.212 | 13.072 | 15.865 | 15.865 | 15.865 |
| 1995 | 0.806 | 1.973 | 4.589 | 7.560 | 9.750 | 11.152 | 13.983 | 14.147 | 14.147 | 14.147 |
| 1996 | 0.787 | 1.877 | 4.639 | 6.997 | 9.854 | 11.407 | 13.040 | 10.363 | 10.363 | 10.363 |
| 1997 | 0.771 | 2.039 | 4.516 | 7.389 | 9.719 | 11.820 | 14.367 | 13.687 | 13.687 | 13.687 |
| 1998 | 0.853 | 1.896 | 4.461 | 6.881 | 9.329 | 11.216 | 13.904 | 14.573 | 17.161 | 14.020 |
| 1999 | 0.993 | 2.098 | 4.495 | 7.326 | 8.945 | 11.255 | 13.877 | 15.988 | 15.988 | 17.159 |


| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 | Age 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 0.863 | 2.541 | 4.629 | 7.042 | 9.502 | 10.660 | 11.746 | 14.476 | 14.720 | 14.720 |
| 2001 | 0.794 | 2.029 | 5.112 | 7.858 | 9.832 | 11.423 | 13.206 | 14.879 | 16.311 | 16.311 |
| 2002 | 0.757 | 1.880 | 4.728 | 6.764 | 9.360 | 10.774 | 12.876 | 13.463 | 13.719 | 14.300 |
| 2003 | 0.889 | 1.844 | 4.274 | 6.667 | 9.506 | 11.064 | 12.040 | 12.762 | 11.139 | 11.139 |
| 2004 | 0.884 | 2.177 | 4.543 | 7.073 | 9.435 | 10.802 | 11.985 | 14.115 | 14.115 | 12.468 |
| 2005 | 0.776 | 2.118 | 3.907 | 6.168 | 9.194 | 11.544 | 10.037 | 12.657 | 13.835 | 13.835 |
| 2006 | 0.789 | 1.793 | 4.716 | 7.404 | 9.186 | 11.646 | 12.313 | 12.699 | 12.699 | 12.699 |
| 2007 | 0.772 | 1.657 | 4.276 | 7.463 | 9.697 | 11.863 | 12.441 | 13.953 | 15.046 | 15.046 |
| 2008 | 0.847 | 1.804 | 4.541 | 7.164 | 9.229 | 11.095 | 13.470 | 12.807 | 15.178 | 16.086 |
| 2009 | 0.923 | 2.384 | 4.248 | 6.721 | 8.895 | 10.584 | 10.342 | 10.497 | 16.169 | 14.560 |
| 2010 | 0.853 | 2.226 | 4.789 | 7.285 | 9.975 | 11.948 | 12.188 | 14.489 | 15.119 | 15.119 |
| 2011 | 0.532 | 1.449 | 4.551 | 7.745 | 9.524 | 10.597 | 12.749 | 10.595 | 10.595 | 10.595 |
| 2012 | 1.093 | 1.712 | 3.510 | 7.077 | 10.196 | 12.232 | 14.106 | 13.929 | 11.214 | 16.248 |
| 2013 | 0.982 | 2.159 | 4.087 | 6.977 | 8.363 | 10.479 | 11.904 | 16.384 | 12.989 | 12.989 |
| 2014 | 0.811 | 2.454 | 4.726 | 7.228 | 9.114 | 11.080 | 12.014 | 16.659 | 16.659 | 16.659 |
| 2015 | 0.915 | 1.838 | 4.144 | 7.980 | 9.539 | 10.719 | 11.891 | 12.416 | 16.165 | 16.165 |
| 2016 | 0.850 | 1.991 | 4.367 | 7.167 | 9.198 | 11.131 | 10.912 | 14.379 | 17.083 | 17.083 |
| 2017 | 0.966 | 2.160 | 3.991 | 7.057 | 8.716 | 9.276 | 10.518 | 11.236 | 12.279 | 12.279 |
| 2018 | 0.909 | 2.143 | 4.334 | 6.795 | 9.180 | 10.006 | 12.565 | 13.676 | 13.121 | 14.726 |

Table 8.4. Cod in divisions 7e-k. Stock weight-at-age =1st quarter values.

| year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 | Age 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0.662 | 1.709 | 4.444 | 7.321 | 9.529 | 11.605 | 13.513 | 15.327 | 15.744 | 15.744 |
| 1972 | 0.662 | 1.709 | 4.444 | 7.321 | 9.529 | 11.605 | 13.513 | 15.327 | 15.744 | 15.744 |
| 1973 | 0.662 | 1.709 | 4.444 | 7.321 | 9.529 | 11.605 | 13.513 | 15.327 | 15.744 | 15.744 |
| 1974 | 0.662 | 1.709 | 4.444 | 7.321 | 9.529 | 11.605 | 13.513 | 15.327 | 15.744 | 15.744 |
| 1975 | 0.662 | 1.709 | 4.444 | 7.321 | 9.529 | 11.605 | 13.513 | 15.327 | 15.744 | 15.744 |
| 1976 | 0.662 | 1.709 | 4.444 | 7.321 | 9.529 | 11.605 | 13.513 | 15.327 | 15.744 | 15.744 |
| 1977 | 0.662 | 1.709 | 4.444 | 7.321 | 9.529 | 11.605 | 13.513 | 15.327 | 15.744 | 15.744 |
| 1978 | 0.662 | 1.709 | 4.444 | 7.321 | 9.529 | 11.605 | 13.513 | 15.327 | 15.744 | 15.744 |
| 1979 | 0.662 | 1.709 | 4.444 | 7.321 | 9.529 | 11.605 | 13.513 | 15.327 | 15.744 | 15.744 |
| 1980 | 0.662 | 1.709 | 4.444 | 7.321 | 9.529 | 11.605 | 13.513 | 15.327 | 15.744 | 15.744 |
| 1981 | 0.460 | 1.549 | 2.284 | 7.806 | 10.544 | 11.439 | 14.464 | 15.354 | 15.354 | 15.354 |
| 1982 | 0.704 | 1.488 | 3.876 | 7.407 | 9.624 | 12.316 | 15.032 | 18.569 | 18.569 | 18.569 |
| 1983 | 0.446 | 1.945 | 4.467 | 7.353 | 9.752 | 11.223 | 15.908 | 18.089 | 21.977 | 21.977 |
| 1984 | 0.512 | 1.951 | 4.928 | 7.433 | 9.552 | 12.180 | 14.181 | 16.733 | 16.733 | 16.733 |
| 1985 | 0.581 | 2.070 | 5.333 | 8.376 | 10.851 | 11.585 | 14.247 | 16.399 | 20.853 | 20.853 |
| 1986 | 0.528 | 1.902 | 5.286 | 7.382 | 10.689 | 12.393 | 14.482 | 14.482 | 14.482 | 14.482 |
| 1987 | 0.522 | 1.947 | 4.877 | 7.946 | 10.308 | 14.419 | 15.171 | 16.201 | 16.201 | 16.201 |
| 1988 | 0.906 | 1.621 | 4.887 | 7.777 | 10.302 | 11.786 | 12.416 | 13.889 | 15.119 | 15.119 |
| 1989 | 0.844 | 1.463 | 4.514 | 7.615 | 9.438 | 12.692 | 12.788 | 17.794 | 17.794 | 17.794 |
| 1990 | 0.613 | 1.774 | 4.390 | 7.186 | 8.486 | 10.703 | 13.305 | 16.987 | 13.081 | 13.081 |
| 1991 | 0.539 | 1.538 | 4.791 | 6.524 | 8.631 | 10.672 | 13.512 | 14.898 | 14.898 | 14.898 |
| 1992 | 0.663 | 1.318 | 4.600 | 6.558 | 9.342 | 11.285 | 12.322 | 14.770 | 14.770 | 14.770 |
| 1993 | 0.703 | 1.385 | 4.278 | 6.574 | 8.066 | 10.815 | 11.945 | 13.421 | 16.530 | 16.530 |
| 1994 | 0.605 | 1.754 | 4.189 | 7.720 | 9.722 | 12.101 | 12.844 | 15.859 | 15.859 | 15.859 |
| 1995 | 0.612 | 1.444 | 4.346 | 7.452 | 9.140 | 10.646 | 13.908 | 14.147 | 14.147 | 14.147 |
| 1996 | 0.673 | 1.283 | 4.471 | 6.747 | 9.877 | 11.424 | 12.848 | 12.848 | 12.848 | 12.848 |
| 1997 | 0.470 | 1.410 | 4.079 | 7.112 | 9.044 | 11.156 | 13.730 | 13.623 | 13.623 | 13.623 |
| 1998 | 0.421 | 1.314 | 4.340 | 6.676 | 9.303 | 11.172 | 12.369 | 14.205 | 17.161 | 14.020 |
| 1999 | 0.778 | 1.542 | 4.252 | 7.126 | 8.700 | 11.142 | 13.978 | 17.463 | 17.159 | 17.159 |


| year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 | Age 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 0.561 | 1.696 | 4.223 | 6.627 | 9.326 | 10.505 | 11.115 | 13.566 | 13.566 | 13.566 |
| 2001 | 0.630 | 1.455 | 4.904 | 7.872 | 10.192 | 11.613 | 13.174 | 14.715 | 16.311 | 16.311 |
| 2002 | 0.352 | 1.257 | 4.452 | 7.046 | 9.400 | 10.614 | 12.637 | 14.949 | 14.949 | 14.949 |
| 2003 | 0.482 | 1.327 | 4.111 | 6.601 | 9.183 | 10.635 | 12.047 | 15.832 | 15.832 | 15.832 |
| 2004 | 0.591 | 1.258 | 4.053 | 6.759 | 9.372 | 10.158 | 11.680 | 13.850 | 13.850 | 13.850 |
| 2005 | 0.588 | 1.688 | 4.075 | 5.945 | 9.018 | 11.333 | 11.487 | 13.772 | 13.772 | 13.772 |
| 2006 | 0.703 | 1.216 | 4.233 | 6.819 | 8.895 | 11.487 | 11.411 | 12.703 | 12.703 | 12.703 |
| 2007 | 0.722 | 1.399 | 3.794 | 6.990 | 9.809 | 12.273 | 15.042 | 14.465 | 14.795 | 14.795 |
| 2008 | 0.869 | 1.449 | 4.188 | 6.896 | 8.881 | 11.543 | 13.624 | 10.045 | 13.763 | 13.763 |
| 2009 | 0.938 | 1.629 | 3.865 | 6.557 | 8.985 | 10.567 | 12.981 | 12.981 | 12.981 | 12.981 |
| 2010 | 0.819 | 1.424 | 4.373 | 6.984 | 9.891 | 11.663 | 12.575 | 13.085 | 13.085 | 13.085 |
| 2011 | 0.374 | 1.214 | 4.198 | 7.239 | 9.404 | 11.039 | 12.785 | 12.785 | 12.785 | 12.785 |
| 2012 | 1.005 | 1.224 | 3.534 | 7.333 | 10.404 | 11.702 | 13.727 | 12.663 | 16.045 | 16.174 |
| 2013 | 0.497 | 1.377 | 3.747 | 6.805 | 8.491 | 9.945 | 9.897 | 17.158 | 17.158 | 17.158 |
| 2014 | 0.464 | 1.654 | 3.788 | 6.530 | 9.074 | 10.584 | 11.611 | 12.285 | 12.285 | 12.285 |
| 2015 | 1.161 | 1.309 | 4.079 | 8.517 | 10.105 | 10.661 | 12.288 | 13.134 | 13.134 | 13.134 |
| 2016 | 0.647 | 1.310 | 3.683 | 6.700 | 10.573 | 11.453 | 12.928 | 16.875 | 16.435 | 16.435 |
| 2017 | 0.299 | 1.543 | 3.363 | 6.191 | 6.376 | 10.606 | 12.327 | 12.793 | 17.514 | 17.514 |
| 2018 | 0.539 | 1.356 | 3.706 | 5.465 | 9.146 | 10.237 | 12.432 | 14.449 | 16.486 | 16.486 |

Table 8.5a. Cod in divisions 7e-k. Time-series of landings, effort and lpue for French OT-DEF fleets. Units: landings in tonnes, effort in 000s hours fished and lpue in kg/hour fished.

| Year | Effort | Landings | Lpue |
| :---: | :---: | :---: | :---: |
| 2000 | 217480.1 | 1360798.3 | 6.26 |
| 2001 | 223428.0 | 2297415.3 | 10.28 |
| 2002 | 191161.1 | 2521943.2 | 13.19 |
| 2003 | 184878.5 | 1594331.4 | 8.62 |
| 2004 | 164606.5 | 693554.3 | 4.21 |
| 2005 | 132471.5 | 589933.2 | 4.45 |
| 2006 | 117258.8 | 571191.5 | 4.87 |
| 2007 | 115878.4 | 816210.8 | 7.04 |
| 2008 | 113485.2 | 652235.7 | 5.75 |
| 2009 | 113347.6 | 550405.7 | 4.86 |
| 2010 | 100331.9 | 635001.8 | 6.33 |
| 2011 | 101251.0 | 925372.7 | 9.14 |
| 2012 | 124404.4 | 2518809.6 | 20.25 |
| 2013 | 155301.2 | 1513472.3 | 9.75 |
| 2014 | 147142.9 | 1097602.2 | 7.46 |
| 2015 | 135732.0 | 1202081.0 | 8.86 |
| 2016 | 131254.0 | 964207.0 | 7.35 |
| 2017 | 130855.0 | 475897.0 | 3.64 |
| 2018 | 112660.0 | 154074.2 | 1.37 |

Table 8.5b. Cod in divisions 7e-k. Time-series of landings, effort and lpue for the Irish fleets. Units: landings in tonnes live weight, effort in 000s hours fished and lpue in $\mathrm{kg} /$ hour fished.

|  | Otter_trawl_27.7j |  |  | Beam_trawl_27.7j |  |  | Scottish_seiner_27.7j |  |  | Gillnet_27.7j |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| X | Landing | Effort | Ipue | Landings | Effort | Ipue | Landings | Effort | Ipue | Landings | Effort | Ipue |
| 1995 | 339.3 | 93.2 | 3.6 | 0.0 | 0.2 | 0.2 | 75.5 | 5.3 | 14.4 | 178.8 | 21.3 | 8.4 |
| 1996 | 326.4 | 70.2 | 4.6 | 8.7 | 1.4 | 6.3 | 124.5 | 8.2 | 15.3 | 65.0 | 5.2 | 12.4 |
| 1997 | 352.7 | 82.7 | 4.3 | 3.4 | 1.7 | 2.0 | 115.8 | 10.7 | 10.8 | 45.5 | 8.3 | 5.5 |
| 1998 | 262.7 | 89.1 | 2.9 | 19.1 | 5.2 | 3.7 | 103.4 | 6.6 | 15.6 | 59.1 | 16.0 | 3.7 |
| 1999 | 76.7 | 40.5 | 1.9 | 27.5 | 7.4 | 3.7 | 9.6 | 1.4 | 6.8 | 24.6 | 8.7 | 2.8 |
| 2000 | 95.5 | 63.9 | 1.5 | 21.2 | 6.9 | 3.1 | 24.4 | 3.5 | 7.0 | 13.8 | 7.0 | 2.0 |
| 2001 | 148.5 | 67.4 | 2.2 | 10.7 | 3.0 | 3.6 | 31.3 | 4.4 | 7.1 | 14.8 | 6.6 | 2.3 |
| 2002 | 150.0 | 90.4 | 1.7 | 5.4 | 3.1 | 1.7 | 24.6 | 8.9 | 2.8 | 12.3 | 8.1 | 1.5 |
| 2003 | 73.6 | 107.4 | 0.7 | 8.8 | 9.0 | 1.0 | 12.0 | 7.9 | 1.5 | 6.3 | 11.2 | 0.6 |
| 2004 | 36.1 | 88.3 | 0.4 | 2.5 | 2.2 | 1.2 | 10.3 | 8.1 | 1.3 | 4.2 | 6.1 | 0.7 |
| 2005 | 37.8 | 71.3 | 0.5 | 4.7 | 2.4 | 2.0 | 17.5 | 5.8 | 3.0 | 3.4 | 6.1 | 0.6 |
| 2006 | 39.6 | 64.5 | 0.6 | 2.0 | 1.5 | 1.3 | 15.6 | 5.3 | 2.9 | 7.2 | 7.3 | 1.0 |
| 2007 | 35.9 | 78.3 | 0.5 | 7.8 | 2.4 | 3.3 | 9.8 | 3.5 | 2.8 | 6.5 | 10.5 | 0.6 |
| 2008 | 33.1 | 66.7 | 0.5 | 2.6 | 1.1 | 2.3 | 9.5 | 2.8 | 3.3 | 6.5 | 7.9 | 0.8 |
| 2009 | 26.6 | 73.0 | 0.4 | 4.7 | 2.8 | 1.7 | 8.9 | 3.3 | 2.7 | 8.0 | 10.9 | 0.7 |
| 2010 | 52.5 | 85.7 | 0.6 | 1.7 | 1.0 | 1.7 | 17.0 | 4.4 | 3.9 | 8.4 | 9.4 | 0.9 |
| 2011 | 57.7 | 62.8 | 0.9 | 1.7 | 0.6 | 2.7 | 21.6 | 4.6 | 4.7 | 16.8 | 8.0 | 2.1 |
| 2012 | 62.8 | 65.6 | 1.0 | 0.4 | 0.3 | 1.5 | 29.8 | 5.4 | 5.6 | 25.2 | 8.3 | 3.0 |
| 2013 | 66.1 | 61.3 | 1.1 | 1.8 | 0.6 | 3.3 | 32.5 | 6.6 | 4.9 | 15.4 | 9.8 | 1.6 |
| 2014 | 51.6 | 53.9 | 1.0 | 1.2 | 0.6 | 1.9 | 52.6 | 7.4 | 7.1 | 9.7 | 12.2 | 0.8 |
| 2015 | 63.6 | 46.9 | 1.4 | 0.6 | 0.1 | 6.3 | 38.2 | 5.3 | 7.2 | 18.1 | 14.2 | 1.3 |
| 2016 | 48.5 | 50.7 | 1.0 | 0.3 | 0.2 | 1.5 | 25.2 | 5.3 | 4.7 | 15.8 | 17.1 | 0.9 |
| 2017 | 41.3 | 56.4 | 0.7 | 0.0 | 0.0 | 10.0 | 24.0 | 5.3 | 4.5 | 10.4 | 18.0 | 0.6 |
| 2018 | 42.3 | 52.1 | 0.8 | 0.2 | 0.1 | 2.4 | 28.5 | 6.4 | 4.5 | 5.9 | 16.8 | 0.4 |


|  | Otter_trawl_27.7g |  |  | Beam_trawl_27.7g |  |  | Scottish_seiner_27.7g |  |  | Gillnet_27.7g |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Landing | Effort | Ipue | Landings | Effort | Ipue | Landings | Effort | Ipue | Landings | Effort | Ipue |
| 1995 | 429.8 | 63.3 | 6.8 | 85.8 | 20.7 | 4.1 | 111.3 | 6.4 | 17.3 | 111.3 | 6.3 | 17.5 |
| 1996 | 569.2 | 60.0 | 9.5 | 112.5 | 26.7 | 4.2 | 164.9 | 9.7 | 16.9 | 164.9 | 6.2 | 26.7 |
| 1997 | 401.9 | 65.0 | 6.2 | 131.5 | 28.1 | 4.7 | 215.2 | 16.1 | 13.4 | 215.2 | 1.9 | 112.7 |
| 1998 | 450.5 | 72.3 | 6.2 | 166.8 | 35.2 | 4.7 | 264.1 | 14.9 | 17.7 | 264.1 | 3.4 | 76.8 |
| 1999 | 300.7 | 51.5 | 5.8 | 190.6 | 40.8 | 4.7 | 64.6 | 8.0 | 8.1 | 64.6 | 8.4 | 7.7 |
| 2000 | 279.4 | 60.6 | 4.6 | 180.6 | 36.8 | 4.9 | 106.0 | 9.9 | 10.8 | 106.0 | 10.1 | 10.5 |
| 2001 | 358.5 | 69.4 | 5.2 | 101.2 | 39.5 | 2.6 | 115.0 | 16.3 | 7.0 | 115.0 | 8.8 | 13.1 |
| 2002 | 212.9 | 77.2 | 2.8 | 57.9 | 31.5 | 1.8 | 71.0 | 20.9 | 3.4 | 71.0 | 6.4 | 11.0 |
| 2003 | 167.2 | 86.8 | 1.9 | 56.8 | 49.2 | 1.2 | 35.6 | 20.1 | 1.8 | 35.6 | 11.1 | 3.2 |
| 2004 | 190.2 | 97.1 | 2.0 | 74.3 | 54.9 | 1.4 | 54.4 | 18.4 | 3.0 | 54.4 | 13.5 | 4.0 |
| 2005 | 292.5 | 124.7 | 2.3 | 118.9 | 49.6 | 2.4 | 64.4 | 14.6 | 4.4 | 64.4 | 10.9 | 5.9 |
| 2006 | 379.4 | 118.0 | 3.2 | 128.6 | 60.5 | 2.1 | 91.0 | 14.8 | 6.2 | 91.0 | 7.8 | 11.6 |
| 2007 | 316.1 | 135.4 | 2.3 | 96.2 | 55.8 | 1.7 | 58.5 | 15.8 | 3.7 | 58.5 | 9.4 | 6.2 |
| 2008 | 344.9 | 125.4 | 2.7 | 85.4 | 37.2 | 2.3 | 55.6 | 11.6 | 4.8 | 55.6 | 14.1 | 3.9 |
| 2009 | 405.9 | 137.1 | 3.0 | 74.4 | 37.9 | 2.0 | 34.6 | 8.2 | 4.2 | 34.6 | 13.8 | 2.5 |
| 2010 | 524.8 | 140.8 | 3.7 | 94.7 | 40.2 | 2.4 | 54.3 | 9.7 | 5.6 | 54.3 | 14.0 | 3.9 |
| 2011 | 438.4 | 120.3 | 3.6 | 82.5 | 35.3 | 2.3 | 46.7 | 11.0 | 4.2 | 46.7 | 11.3 | 4.1 |
| 2012 | 780.7 | 127.7 | 6.1 | 161.9 | 40.3 | 4.0 | 111.5 | 14.1 | 7.9 | 111.5 | 15.4 | 7.2 |
| 2013 | 721.4 | 118.2 | 6.1 | 195.8 | 38.5 | 5.1 | 111.3 | 13.2 | 8.5 | 111.3 | 14.4 | 7.7 |
| 2014 | 600.1 | 127.3 | 4.7 | 142.9 | 37.8 | 3.8 | 110.5 | 12.5 | 8.9 | 110.5 | 14.1 | 7.8 |
| 2015 | 526.3 | 132.7 | 4.0 | 160.1 | 37.8 | 4.2 | 59.2 | 9.3 | 6.4 | 59.2 | 12.5 | 4.7 |
| 2016 | 418.1 | 148.2 | 2.8 | 106.8 | 39.6 | 2.7 | 51.1 | 10.4 | 4.9 | 51.1 | 13.6 | 3.8 |
| 2017 | 361.4 | 136.1 | 2.7 | 46.4 | 35.2 | 1.3 | 42.1 | 9.7 | 4.3 | 42.1 | 14.8 | 2.8 |
| 2018 | 387.6 | 108.2 | 3.6 | 72.6 | 37.4 | 1.9 | 61.1 | 9.7 | 6.3 | 61.1 | 14.0 | 4.4 |

Table 8.5c. Cod in divisions 7e-k. Time-series of landings, effort and lpue for the UK fleets. Units: landings in tonnes, effort in days fished and lpue in $\mathrm{kg} /$ day.

|  | Beam_trawl_27.7ek |  |  | Trawl_27.7ek |  |  | Trawl_27.7e |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Lpue | Landing | Effort | Lpue | Landings | Effort | Lpue | Landings | Effort |
| 1983 | 8.96 | 25.55 | 2853 | 15.91 | 40.93 | 2573 | 11.01 | 20.6 | 1871 |
| 1984 | 15.28 | 128.75 | 8427 | 29.12 | 235.68 | 8092 | 13.6 | 76.42 | 5618 |
| 1985 | 18.87 | 145.39 | 7706 | 34.88 | 250.67 | 7186 | 11.82 | 63.97 | 5411 |
| 1986 | 24.92 | 165.76 | 6651 | 37.61 | 232.19 | 6174 | 17.7 | 78.31 | 4425 |
| 1987 | 30.88 | 248.91 | 8060 | 38.63 | 210.36 | 5446 | 23.91 | 88.49 | 3701 |
| 1988 | 26.27 | 249.21 | 9487 | 46.53 | 262.68 | 5645 | 35.49 | 151.35 | 4265 |
| 1989 | 22.96 | 231.24 | 10071 | 29.53 | 177.12 | 5997 | 20.84 | 96 | 4607 |
| 1990 | 29.5 | 309.07 | 10477 | 45.91 | 305.78 | 6661 | 27 | 119.41 | 4423 |
| 1991 | 28.41 | 256.19 | 9017 | 40.81 | 242.33 | 5938 | 20.88 | 83.6 | 4004 |
| 1992 | 31.32 | 256.33 | 8183 | 35.7 | 231.85 | 6494 | 19.66 | 80.76 | 4108 |
| 1993 | 23.32 | 221.79 | 9511 | 36.21 | 183.05 | 5055 | 11.4 | 42.88 | 3761 |
| 1994 | 12.86 | 179.13 | 13925 | 17.67 | 78.23 | 4426 | 12.05 | 41.25 | 3423 |
| 1995 | 16.01 | 241.35 | 15076 | 26.12 | 115.05 | 4405 | 16.72 | 55.09 | 3294 |
| 1996 | 19.32 | 304.22 | 15748 | 26.91 | 120.46 | 4476 | 22.87 | 59.21 | 2589 |
| 1997 | 18.55 | 303.67 | 16373 | 29.48 | 150.01 | 5088 | 26.51 | 79.81 | 3011 |
| 1998 | 17.09 | 266.15 | 15574 | 25.28 | 119.56 | 4729 | 23.16 | 62.5 | 2699 |
| 1999 | 16.49 | 257.43 | 15614 | 13.66 | 90.68 | 6638 | 18.83 | 46.81 | 2486 |
| 2000 | 11.43 | 188.07 | 16456 | 15.71 | 110.79 | 7054 | 19.62 | 52.59 | 2681 |
| 2001 | 14.84 | 257.24 | 17335 | 18.68 | 109.75 | 5875 | 21.62 | 59.05 | 2732 |
| 2002 | 8.01 | 132.13 | 16503 | 14.62 | 82.7 | 5657 | 13.93 | 34.11 | 2448 |
| 2003 | 5.95 | 108.77 | 18285 | 11.48 | 58.8 | 5120 | 10.77 | 24.48 | 2273 |
| 2004 | 5.31 | 96.93 | 18250 | 8.36 | 44.06 | 5273 | 6.45 | 15.05 | 2334 |
| 2005 | 6.04 | 103.6 | 17157 | 8.15 | 41.13 | 5047 | 9.86 | 17.38 | 1762 |
| 2006 | 5.96 | 91.88 | 15412 | 10.43 | 55.43 | 5314 | 7.97 | 13.54 | 1699 |
| 2007 | 7.38 | 111.28 | 15085 | 8.74 | 49.65 | 5679 | 11.27 | 21.61 | 1917 |
| 2008 | 5.2 | 71.38 | 13734 | 10.53 | 49.34 | 4686 | 13.86 | 24.26 | 1750 |
| 2009 | 5.53 | 67.27 | 12170 | 5.59 | 27.56 | 4928 | 6.8 | 12.56 | 1847 |
| 2010 | 5.4 | 65.62 | 12150 | 6 | 31.13 | 5185 | 6.9 | 15.27 | 2213 |
| 2011 | 7.5 | 99.03 | 13205 | 10.96 | 47.73 | 4354 | 13.47 | 26 | 1931 |
| 2012 | 12.35 | 165.63 | 13411 | 18.33 | 79.03 | 4312 | 14.96 | 30.95 | 2068 |
| 2013 | 8.84 | 114.49 | 12950 | 18.52 | 37.3 | 2014 | 14.46 | 22.94 | 1587 |
| 2014 | 6.84 | 87.55 | 12807 | 10.63 | 17.07 | 1606 | 9.76 | 14.06 | 1440 |
| 2015 | 7 | 89.39 | 12769 | 15.72 | 16.68 | 1061 | 14.73 | 14.4 | 978 |
| 2016 | 5.3 | 73.81 | 13913 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 2.49 | 35.49 | 14280 | 5.22 | 19.37 | 3711 | 3.9 | 9.33 | 2392 |
| 2018 | 1.87 | 24.41 | 13065 | 5.45 | 17.5 | 3214 | 2.71 | 5.33 | 1967 |

Table 8.6. Cod in divisions 7e-k. Time-series of survey indices scrutinized at WGCSE and used in the assessment.

## Cod in divisions 7e-k, tuning fleets, WGCSE18

## 102

| FR-OTDEF Q2+3+4 trawlers in 7e-k |  |  |  |
| :---: | :---: | :---: | :---: |
| 2000 | 2018 | 0.25 | 1 |
| 1 | 1 | 10 |  |
| 1 |  |  |  |


| Year | Effort | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 | Age 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 217479 | 200742 | 93804 | 59384 | 35784 | 11253 | 5683 | 3988 | 545 | 356 | 0 |
| 2001 | 223427 | 119879 | 383175 | 45401 | 44844 | 34907 | 11427 | 5256 | 2109 | 0 | 0 |
| 2002 | 191161 | 188306 | 472476 | 144332 | 38748 | 16046 | 9760 | 4317 | 4212 | 252 | 0 |
| 2003 | 184878 | 22380 | 134512 | 138065 | 59698 | 7928 | 7313 | 4455 | 847 | 424 | 0 |
| 2004 | 164606 | 12412 | 54908 | 41644 | 21032 | 13420 | 1720 | 208 | 0 | 0 | 208 |
| 2005 | 132472 | 13489 | 132632 | 10525 | 6207 | 8814 | 2861 | 367 | 54 | 237 | 0 |
| 2006 | 117259 | 24447 | 148506 | 27730 | 3716 | 1912 | 1282 | 845 | 0 | 0 | 0 |
| 2007 | 115878 | 265362 | 409573 | 76766 | 13367 | 2099 | 684 | 818 | 235 | 60 | 0 |
| 2008 | 113485 | 77385 | 252690 | 44372 | 16057 | 4178 | 624 | 236 | 447 | 0 | 8 |
| 2009 | 113348 | 106600 | 58211 | 46807 | 14017 | 5042 | 1939 | 894 | 353 | 0 | 19 |
| 2010 | 100332 | 206831 | 103580 | 15881 | 8766 | 4600 | 678 | 102 | 0 | 17 | 0 |
| 2011 | 101251 | 6870 | 1145981 | 92577 | 22801 | 17131 | 3074 | 551 | 0 | 0 | 0 |
| 2012 | 124404 | 2709 | 108920 | 463339 | 109825 | 12257 | 6173 | 1939 | 176 | 1329 | 0 |
| 2013 | 155301 | 41174 | 66032 | 126952 | 129554 | 21809 | 5676 | 1921 | 0 | 0 | 0 |
| 2014 | 147143 | 160520 | 70506 | 23843 | 29394 | 48405 | 2958 | 191 | 0 | 0 | 0 |
| 2015 | 135732 | 3473 | 409342 | 36700 | 6263 | 11629 | 7460 | 4640 | 0 | 0 | 0 |
| 2016 | 131254 | 11768 | 21661 | 149990 | 12802 | 2733 | 2975 | 6765 | 0 | 0 | 0 |
| 2017 | 130855 | 2135 | 83620 | 19322 | 23688 | 1831 | 946 | 1226 | 0 | 0 | 0 |
| 2018 | 112660 | 11660 | 14429 | 10793 | 1333 | 4676 | 365 | 198 | 0 | 0 | 0 |

## IR-GFS FR-EVHOE Q4 combined indices new

| 2003 | 2018 |  |  |
| :---: | :---: | :---: | :---: |
| 1 | 1 | 0.79 | 0.92 |
| 0 | 6 |  |  |


| Year | Effort | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 1 | 0.14 | 0.61 | 0.75 | 0.50 | 0.17 | 0.00 | 0.00 |
| 2004 | 1 | 0.24 | 0.88 | 0.24 | 0.15 | 0.14 | 0.07 | 0.00 |
| 2005 | 1 | 0.06 | 1.81 | 0.26 | 0.09 | 0.00 | 0.00 | 0.00 |
| 2006 | 1 | 0.04 | 1.39 | 0.67 | 0.08 | 0.00 | 0.00 | 0.02 |
| 2007 | 1 | 0.00 | 1.93 | 0.64 | 0.19 | 0.05 | 0.00 | 0.00 |
| 2008 | 1 | 0.00 | 0.55 | 0.88 | 0.24 | 0.12 | 0.00 | 0.00 |
| 2009 | 1 | 0.10 | 1.38 | 0.17 | 0.26 | 0.12 | 0.00 | 0.01 |
| 2010 | 1 | 0.12 | 7.34 | 0.76 | 0.04 | 0.06 | 0.07 | 0.00 |
| 2011 | 1 | 0.02 | 4.09 | 3.54 | 0.22 | 0.04 | 0.03 | 0.00 |
| 2012 | 1 | 0.00 | 0.39 | 1.32 | 0.80 | 0.19 | 0.04 | 0.00 |
| 2013 | 1 | 0.08 | 0.42 | 0.05 | 0.21 | 0.23 | 0.00 | 0.00 |
| 2014 | 1 | 0.00 | 3.64 | 0.27 | 0.12 | 0.15 | 0.20 | 0.00 |
| 2015 | 1 | 0.00 | 0.31 | 1.36 | 0.12 | 0.00 | 0.05 | 0.06 |
| 2016 | 1 | 0.00 | 2.27 | 0.18 | 0.81 | 0.07 | 0.02 | 0.07 |
| 2017 | 1 | 0.04 | 0.32 | 0.60 | 0.27 | 0.17 | 0.02 | 0.00 |
| 2018 | 1 | 0.05 | 0.48 | 0.09 | 0.04 | 0.04 | 0.12 | 0.00 |

Table 8.7. Cod in divisions 7e-k. Final XSA diagnostics (from FLR XSA).
FLR XSA Diagnostics 2019-05-10 12:51:38

Cpue data from indices

Catch data for 48 years. 1971 to 2018. Ages 1 to 7.


Fishing mortalities

| age | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.072 | 0.052 | 0.513 | 0.057 | 0.113 | 0.167 | 0.089 | 0.116 | 0.048 | 0.117 |
| 2 | 0.493 | 0.376 | 0.681 | 0.411 | 0.714 | 0.782 | 0.851 | 0.428 | 0.882 | 0.966 |
| 3 | 0.786 | 0.493 | 0.416 | 0.877 | 0.992 | 0.918 | 0.9 | 1.172 | 1.146 | 1.002 |
| 4 | 0.692 | 0.596 | 0.446 | 1.048 | 0.985 | 0.861 | 0.701 | 0.775 | 0.909 | 0.547 |
| 5 | 1.174 | 0.837 | 0.461 | 0.454 | 1.242 | 0.998 | 0.74 | 0.471 | 1.27 | 0.819 |
| 6 | 0.831 | 1.185 | 0.439 | 0.395 | 0.89 | 0.552 | 0.564 | 0.527 | 0.691 | 0.973 |
| 7 | 0.831 | 1.185 | 0.439 | 0.395 | 0.89 | 0.552 | 0.564 | 0.527 | 0.691 | 0.973 |

XSA population number (NA)

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 2995 | 867 | 692 | 220 | 61 | 20 | 4 |
| 2010 | 13499 | 1670 | 366 | 233 | 84 | 15 | 3 |
| 2011 | 4874 | 7677 | 794 | 165 | 98 | 28 | 6 |
| 2012 | 813 | 1748 | 2688 | 387 | 81 | 48 | 24 |
| 2013 | 1327 | 460 | 802 | 825 | 104 | 40 | 11 |
| 2014 | 6277 | 710 | 156 | 219 | 235 | 23 | 3 |
| 2015 | 540 | 3184 | 225 | 46 | 71 | 68 | 22 |
| 2016 | 1561 | 296 | 941 | 67 | 17 | 26 | 47 |
| 2017 | 496 | 834 | 134 | 215 | 24 | 8 | 15 |
| 2018 | 990 | 284 | 239 | 31 | 66 | 5 | 3 |

Estimated population abundance at 1st January 2019

| year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2019 | 0 | 528 | 75 | 65 | 14 | 23 | 2 |

Fleet: FR-OTDEF
Log catchability residuals.

|  | year |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
|  | 2006 | 2007 | 2008 | 2009 | 2010 |  |
| 1 | 0.036 | -0.354 | 1.628 | 0.037 | -1.216 | -1.309 |
|  | -0.678 | 1.932 | 1.551 | 1.229 | 0.497 |  |
| 2 | -0.421 | -0.621 | -0.13 | 0.038 | -0.335 | 0.036 |
|  | -0.162 | 0.895 | 0.511 | -0.252 | -0.278 |  |
| 3 | -0.666 | -0.244 | -0.253 | -0.072 | 0.111 | -0.597 |
|  | -0.335 | 0.306 | -0.24 | -0.238 | -0.731 |  |
| 4 | -0.592 | 0.141 | 0.86 | 0.27 | -0.419 | -0.299 |
|  | -0.313 | 0.027 | 0.004 | -0.371 | -0.831 |  |
| 5 | -0.607 | 0.416 | 0.188 | 0.32 | 0.263 | 0.248 |
|  | -0.117 | 0.112 | -0.146 | 0.158 | -0.331 |  |
| 6 | -0.014 | 0.23 | 0.097 | 0.137 | -0.065 | 0.067 |
|  | $-0.182$ <br> year | 0.009 | -0.018 | 0.114 | -0.307 |  |
| age | 2017 | 2018 |  |  |  |  |
| 1 | -1.042 | 0.156 |  |  |  |  |
| 2 | 0.23 | -0.252 |  |  |  |  |
| 3 | 0.583 | -0.512 |  |  |  |  |
| 4 | 0.157 | -0.855 |  |  |  |  |
| 5 | -0.009 | -0.202 |  |  |  |  |
| 6 | 0.021 | -0.13 |  |  |  |  |

Mean $\log$ catchability and standard error of ages with catchability independent of year-class strength and constant w.r.t. time

|  | $\mathbf{1}$ | 2 | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean_Logq | -8.9383 | -6.6585 | -6.5329 | -6.5329 | -6.5329 | -6.5329 |
| S.E_Logq | 0.5722 | 0.5722 | 0.5722 | 0.5722 | 0.5722 | 0.5722 |

Fleet: IR-FR COMBINED SURVEY
Log catchability residuals.

| age | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2009 | 2010 | 2011 | 2012 | 2013 |  |
| 1 | -0.083 | -0.482 | -0.171 | -0.517 | 0.043 | -0.402 |
|  | -0.134 | 0.015 | 0.843 | -0.105 | -0.474 |  |
| 2 | 0.725 | -0.051 | -0.686 | -0.177 | -0.139 | 0.177 |
|  | -0.807 | -0.065 | 0.209 | 0.471 | -1.208 |  |
| 3 | 0.26 | 0.243 | 0.239 | -0.77 | -0.267 | -0.119 |
|  | -0.094 | -1.58 | -0.715 | -0.249 | -0.279 |  |
| 4 | 0.307 | 0.346 | NaN | NaN | -0.238 | 0.467 |
|  | 0.169 | -0.663 | -0.853 | 0.368 | -0.252 |  |
| age | 2014 | 2015 | 2016 | 2017 | 2018 |  |
| 1 | 0.177 | 0.101 | 1.053 | 0.181 | -0.044 |  |
| 2 | 0.102 | 0.277 | 0.269 | 0.826 | 0.078 |  |
| 3 | 0.736 | 0.354 | 1.065 | 1.895 | -0.718 |  |
| 4 | 0.539 | NaN | 0.883 | 0.725 | 0.894 |  |

Mean log catchability and standard error of ages with catchability independent of year-class strength and constant w.r.t. time

|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Mean_Logq | -7.0498 | -6.9942 | -6.8614 | -6.8614 |
| S.E_Logq | 0.595 | 0.595 | 0.595 | 0.595 |

Terminal year survivor and F summaries:
Age 1 Year class = 2017
source

FR-OTDEF Q2 $+3+4$ trawlers in VIle-k IR-GFS FR-EVHOE Q4 combined indices new 505 fshk

Age 2 Year class $=2016$
source

FR-OTDEF Q2+3+4 trawlers in VIIe-k
IR-GFS FR-EVHOE Q4 combined indices new 86 fshk

Age 3 Year class $=2015$
source

FR-OTDEF Q2+3+4 trawlers in VIIe-k 49
IR-GFS FR-EVHOE Q4 combined indices new 100 fshk

Age 4 Year class $=2014$
source

FR-OTDEF Q2+3+4 trawlers in VIIe-k 9
IR-GFS FR-EVHOE Q4 combined indices new 29
fshk

Age 5 Year class $=2013$
source

FR-OTDEF Q2+3+4 trawlers in VIIe-k 21
IR-GFS FR-EVHOE Q4 combined indices new 43
fshk

Age 6 Year class = 2012
source

FR-OTDEF Q2+3+4 trawlers in VIIe-k 1
IR-GFS FR-EVHOE Q4 combined indices new 2
fshk
10.720
10.116

574
2.361

110
0.326
3.481

61
40.384
$4 \quad 0.474$

7

21
40.135
$5 \quad 0.677$
18
$4 \quad 0.035$
$6 \quad 0.807$
survivors N scaledWts
survivors N scaledWts
10.188
survivors N scaledWts
10.163
survivors N scaledWts
10.158
survivors N scaledWts
10.193
survivors N scaledWts
10.142
10.157

Table 8.8. Cod in divisions 7e-k. Final XSA fishing mortality-at-age.

| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ | $\mathrm{F}_{\text {bar }}($ mean 2-5) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0.219 | 0.685 | 0.635 | 0.550 | 0.359 | 0.519 | 0.519 | 0.557 |
| 1972 | 0.006 | 0.518 | 0.397 | 0.567 | 0.585 | 0.521 | 0.521 | 0.517 |
| 1973 | 0.165 | 0.731 | 0.553 | 0.437 | 0.442 | 0.481 | 0.481 | 0.541 |
| 1974 | 0.001 | 0.208 | 0.290 | 0.407 | 0.595 | 0.435 | 0.435 | 0.375 |
| 1975 | 0.156 | 0.367 | 0.313 | 1.160 | 0.871 | 0.791 | 0.791 | 0.678 |
| 1976 | 0.034 | 0.838 | 0.327 | 0.333 | 0.717 | 0.463 | 0.463 | 0.554 |
| 1977 | 0.011 | 0.570 | 0.352 | 0.135 | 0.338 | 0.277 | 0.277 | 0.349 |
| 1978 | 0.097 | 0.431 | 0.405 | 0.322 | 0.287 | 0.340 | 0.340 | 0.361 |
| 1979 | 0.089 | 0.340 | 0.512 | 0.433 | 0.585 | 0.515 | 0.515 | 0.468 |
| 1980 | 0.066 | 0.513 | 0.601 | 0.896 | 0.767 | 0.764 | 0.764 | 0.694 |
| 1981 | 0.082 | 0.776 | 0.987 | 0.890 | 0.580 | 0.829 | 0.829 | 0.808 |
| 1982 | 0.048 | 0.680 | 0.653 | 0.642 | 0.586 | 0.633 | 0.633 | 0.640 |
| 1983 | 0.274 | 0.747 | 0.987 | 0.861 | 0.741 | 0.874 | 0.874 | 0.834 |
| 1984 | 0.153 | 0.617 | 0.499 | 0.514 | 0.496 | 0.507 | 0.507 | 0.532 |
| 1985 | 0.175 | 0.577 | 0.624 | 0.505 | 0.450 | 0.531 | 0.531 | 0.539 |
| 1986 | 0.184 | 0.752 | 0.881 | 0.978 | 0.536 | 0.808 | 0.808 | 0.787 |
| 1987 | 0.148 | 0.596 | 0.757 | 1.165 | 0.726 | 0.894 | 0.894 | 0.811 |
| 1988 | 0.215 | 0.691 | 0.497 | 0.606 | 0.763 | 0.629 | 0.629 | 0.639 |
| 1989 | 0.269 | 0.768 | 1.010 | 0.510 | 0.978 | 0.843 | 0.843 | 0.816 |
| 1990 | 0.121 | 0.923 | 0.942 | 0.954 | 0.602 | 0.843 | 0.843 | 0.855 |
| 1991 | 0.170 | 0.865 | 0.869 | 1.173 | 1.018 | 1.034 | 1.034 | 0.981 |
| 1992 | 0.172 | 0.786 | 0.948 | 0.724 | 0.948 | 0.884 | 0.884 | 0.851 |
| 1993 | 0.101 | 0.656 | 0.908 | 0.705 | 0.726 | 0.789 | 0.789 | 0.749 |
| 1994 | 0.135 | 0.536 | 1.067 | 0.793 | 0.603 | 0.831 | 0.831 | 0.750 |
| 1995 | 0.116 | 0.811 | 0.588 | 0.920 | 0.585 | 0.706 | 0.706 | 0.726 |
| 1996 | 0.114 | 0.743 | 1.001 | 0.618 | 1.059 | 0.904 | 0.904 | 0.855 |
| 1997 | 0.166 | 0.838 | 1.056 | 0.839 | 0.506 | 0.810 | 0.810 | 0.810 |
| 1998 | 0.180 | 0.927 | 0.993 | 1.102 | 0.837 | 0.990 | 0.990 | 0.965 |
| 1999 | 0.319 | 0.795 | 1.009 | 0.817 | 0.927 | 0.930 | 0.930 | 0.887 |


| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ | $F_{\text {bar }}($ mean 2-5) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 0.230 | 0.788 | 0.976 | 0.730 | 0.607 | 0.817 | 0.817 | 0.775 |
| 2001 | 0.174 | 0.825 | 0.676 | 0.723 | 0.858 | 0.612 | 0.612 | 0.770 |
| 2002 | 0.133 | 0.817 | 1.103 | 0.932 | 0.620 | 0.351 | 0.351 | 0.868 |
| 2003 | 0.108 | 0.875 | 1.227 | 1.075 | 0.629 | 0.354 | 0.354 | 0.952 |
| 2004 | 0.163 | 0.721 | 0.960 | 1.061 | 1.143 | 0.903 | 0.903 | 0.971 |
| 2005 | 0.096 | 0.771 | 1.155 | 1.017 | 1.065 | 0.706 | 0.706 | 1.002 |
| 2006 | 0.110 | 0.750 | 0.938 | 0.741 | 0.953 | 1.194 | 1.194 | 0.845 |
| 2007 | 0.180 | 0.943 | 1.031 | 0.692 | 0.718 | 0.866 | 0.866 | 0.846 |
| 2008 | 0.104 | 0.656 | 0.799 | 0.901 | 0.727 | 0.901 | 0.901 | 0.771 |
| 2009 | 0.072 | 0.493 | 0.786 | 0.692 | 1.174 | 0.831 | 0.831 | 0.786 |
| 2010 | 0.052 | 0.376 | 0.493 | 0.596 | 0.837 | 1.185 | 1.185 | 0.576 |
| 2011 | 0.513 | 0.681 | 0.416 | 0.446 | 0.461 | 0.439 | 0.439 | 0.501 |
| 2012 | 0.057 | 0.411 | 0.877 | 1.048 | 0.454 | 0.395 | 0.395 | 0.698 |
| 2013 | 0.113 | 0.714 | 0.992 | 0.985 | 1.242 | 0.890 | 0.890 | 0.983 |
| 2014 | 0.167 | 0.782 | 0.918 | 0.861 | 0.998 | 0.552 | 0.552 | 0.890 |
| 2015 | 0.089 | 0.851 | 0.900 | 0.701 | 0.740 | 0.564 | 0.564 | 0.798 |
| 2016 | 0.116 | 0.428 | 1.172 | 0.775 | 0.471 | 0.527 | 0.527 | 0.712 |
| 2017 | 0.048 | 0.882 | 1.146 | 0.909 | 1.270 | 0.691 | 0.691 | 1.052 |
| 2018 | 0.117 | 0.966 | 1.002 | 0.547 | 0.819 | 0.973 | 0.973 | 0.834 |

Table 8.9. Cod in divisions 7e-k. Final XSA stock number-at-age.

| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 4769 | 1118 | 1381 | 260 | 131 | 47 | 30 |
| 1972 | 928 | 2296 | 390 | 540 | 115 | 72 | 46 |
| 1973 | 2810 | 553 | 947 | 194 | 234 | 50 | 32 |
| 1974 | 889 | 1428 | 184 | 402 | 96 | 118 | 113 |
| 1975 | 6031 | 532 | 802 | 102 | 204 | 41 | 67 |
| 1976 | 1986 | 3093 | 255 | 433 | 24 | 67 | 21 |
| 1977 | 2871 | 1151 | 926 | 136 | 237 | 9 | 78 |
| 1978 | 2741 | 1701 | 450 | 480 | 91 | 132 | 77 |
| 1979 | 6630 | 1491 | 765 | 222 | 266 | 53 | 88 |
| 1980 | 12254 | 3634 | 734 | 338 | 110 | 116 | 29 |
| 1981 | 5179 | 6872 | 1506 | 297 | 105 | 40 | 24 |
| 1982 | 2117 | 2860 | 2189 | 414 | 93 | 46 | 12 |
| 1983 | 6923 | 1209 | 1003 | 841 | 167 | 41 | 21 |
| 1984 | 6696 | 3153 | 396 | 276 | 272 | 62 | 14 |
| 1985 | 5892 | 3443 | 1177 | 178 | 126 | 129 | 41 |
| 1986 | 5000 | 2964 | 1338 | 466 | 82 | 63 | 34 |
| 1987 | 25361 | 2493 | 967 | 409 | 134 | 37 | 28 |
| 1988 | 12239 | 13110 | 950 | 335 | 97 | 51 | 19 |
| 1989 | 3648 | 5919 | 4547 | 427 | 140 | 36 | 21 |
| 1990 | 4042 | 1670 | 1900 | 1221 | 196 | 41 | 31 |
| 1991 | 11365 | 2146 | 459 | 547 | 360 | 84 | 24 |
| 1992 | 11742 | 5745 | 625 | 142 | 129 | 101 | 32 |
| 1993 | 3698 | 5927 | 1812 | 179 | 53 | 39 | 32 |
| 1994 | 13711 | 2004 | 2128 | 539 | 67 | 20 | 33 |
| 1995 | 9666 | 7180 | 811 | 540 | 187 | 29 | 11 |
| 1996 | 7410 | 5156 | 2207 | 332 | 165 | 81 | 7 |
| 1997 | 9985 | 3963 | 1698 | 599 | 137 | 45 | 12 |
| 1998 | 5005 | 5067 | 1187 | 436 | 198 | 64 | 14 |
| 1999 | 2348 | 2504 | 1387 | 324 | 111 | 67 | 26 |


| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 10649 | 1022 | 783 | 373 | 109 | 34 | 28 |
| 2001 | 8837 | 5072 | 322 | 218 | 137 | 47 | 46 |
| 2002 | 2182 | 4452 | 1538 | 121 | 81 | 46 | 41 |
| 2003 | 1298 | 1145 | 1361 | 376 | 36 | 34 | 26 |
| 2004 | 2925 | 698 | 331 | 294 | 98 | 15 | 13 |
| 2005 | 4163 | 1488 | 235 | 93 | 78 | 24 | 4 |
| 2006 | 4571 | 2266 | 477 | 55 | 26 | 21 | 8 |
| 2007 | 3850 | 2455 | 741 | 138 | 20 | 8 | 8 |
| 2008 | 1606 | 1927 | 662 | 195 | 53 | 8 | 7 |
| 2009 | 2995 | 867 | 692 | 220 | 61 | 20 | 4 |
| 2010 | 13499 | 1670 | 366 | 233 | 84 | 15 | 3 |
| 2011 | 4874 | 7677 | 794 | 165 | 98 | 28 | 6 |
| 2012 | 813 | 1748 | 2688 | 387 | 81 | 48 | 24 |
| 2013 | 1327 | 460 | 802 | 825 | 104 | 40 | 11 |
| 2014 | 6277 | 710 | 156 | 219 | 235 | 23 | 3 |
| 2015 | 540 | 3184 | 225 | 46 | 71 | 68 | 22 |
| 2016 | 1561 | 296 | 941 | 67 | 17 | 26 | 47 |
| 2017 | 496 | 834 | 134 | 215 | 24 | 8 | 15 |
| 2018 | 990 | 284 | 239 | 31 | 66 | 5 | 3 |
| GMST_71_2016 | 4171 | 2186 | 808 | 273 | 102 | 41 | 21 |
| AMST_71_2016 | 5780 | 2990 | 1049 | 340 | 124 | 50 | 29 |

Table 8.10a. Cod in divisions 7e-k. Final XSA summary table.

| Year | Recruitment | SSB | Catch | Landings | TSB | $\mathrm{F}_{\text {bar }}$ (2-5) | Y/SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 4769 | 10093 | 5782 | 5782 | 15346 | 0.557 | 0.57 |
| 1972 | 928 | 9298 | 4737 | 4737 | 12808 | 0.517 | 0.51 |
| 1973 | 2810 | 8617 | 4015 | 4015 | 11700 | 0.541 | 0.47 |
| 1974 | 889 | 8327 | 2898 | 2898 | 10717 | 0.375 | 0.35 |
| 1975 | 6031 | 7526 | 3993 | 3993 | 12589 | 0.678 | 0.53 |
| 1976 | 1986 | 7316 | 4818 | 4818 | 12224 | 0.554 | 0.66 |
| 1977 | 2871 | 8841 | 3059 | 3059 | 12545 | 0.349 | 0.35 |
| 1978 | 2741 | 9689 | 3647 | 3647 | 13783 | 0.361 | 0.38 |
| 1979 | 6630 | 9848 | 4650 | 4650 | 16346 | 0.467 | 0.47 |
| 1980 | 12254 | 10347 | 7243 | 7243 | 22845 | 0.694 | 0.70 |
| 1981 | 5179 | 11212 | 10597 | 10597 | 20697 | 0.808 | 0.95 |
| 1982 | 2117 | 13547 | 8766 | 8766 | 18951 | 0.640 | 0.65 |
| 1983 | 6923 | 13008 | 9641 | 9641 | 18545 | 0.834 | 0.74 |
| 1984 | 6696 | 9568 | 6631 | 6631 | 17147 | 0.531 | 0.69 |
| 1985 | 5892 | 13103 | 8317 | 8317 | 21794 | 0.539 | 0.63 |
| 1986 | 5000 | 13692 | 10475 | 10475 | 20931 | 0.787 | 0.77 |
| 1987 | 25361 | 11364 | 10228 | 10228 | 28403 | 0.811 | 0.90 |
| 1988 | 12239 | 16606 | 17191 | 17191 | 41445 | 0.639 | 1.04 |
| 1989 | 3648 | 26324 | 19809 | 19809 | 37580 | 0.817 | 0.75 |
| 1990 | 4042 | 19126 | 12749 | 12749 | 25110 | 0.855 | 0.67 |
| 1991 | 11365 | 10846 | 9336 | 9336 | 19520 | 0.981 | 0.86 |
| 1992 | 11742 | 9073 | 9747 | 9747 | 21916 | 0.851 | 1.07 |
| 1993 | 3698 | 12280 | 10425 | 10425 | 20977 | 0.749 | 0.85 |
| 1994 | 13711 | 14357 | 10620 | 10620 | 26246 | 0.750 | 0.74 |
| 1995 | 9666 | 13019 | 11709 | 11709 | 25999 | 0.726 | 0.90 |
| 1996 | 7410 | 15899 | 12681 | 12681 | 26362 | 0.855 | 0.80 |
| 1997 | 9985 | 14065 | 12035 | 12035 | 23366 | 0.810 | 0.86 |
| 1998 | 5005 | 12522 | 11431 | 11431 | 19564 | 0.965 | 0.91 |
| 1999 | 2348 | 10883 | 8594 | 8594 | 15994 | 0.887 | 0.79 |


| Year | Recruitment | SSB | Catch | Landings | TSB | $\mathrm{F}_{\text {bar }}$ (2-5) | Y/SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 10649 | 7548 | 6536 | 6536 | 15182 | 0.775 | 0.87 |
| 2001 | 8837 | 8421 | 8308 | 8308 | 18815 | 0.771 | 0.99 |
| 2002 | 2182 | 10738 | 9236 | 9236 | 15870 | 0.868 | 0.86 |
| 2003 | 1298 | 8795 | 6420 | 6420 | 11249 | 0.951 | 0.73 |
| 2004 | 2925 | 4596 | 3672 | 3672 | 7174 | 0.971 | 0.80 |
| 2005 | 4163 | 3359 | 3062 | 3062 | 7503 | 1.002 | 0.91 |
| 2006 | 4571 | 3739 | 3776 | 3776 | 8922 | 0.846 | 1.01 |
| 2007 | 3850 | 5084 | 4830 | 4830 | 10391 | 0.846 | 0.95 |
| 2008 | 1606 | 5402 | 3961 | 3961 | 8954 | 0.771 | 0.73 |
| 2009 | 2995 | 5022 | 3292 | 3292 | 9142 | 0.786 | 0.66 |
| 2010 | 13499 | 4875 | 3229 | 3229 | 17703 | 0.575 | 0.66 |
| 2011 | 4874 | 8961 | 7261 | 7261 | 16985 | 0.501 | 0.81 |
| 2012 | 813 | 13492 | 7692 | 7692 | 17047 | 0.698 | 0.57 |
| 2013 | 1327 | 9470 | 6290 | 6290 | 11300 | 0.984 | 0.66 |
| 2014 | 6277 | 4716 | 3879 | 3879 | 8523 | 0.890 | 0.82 |
| 2015 | 540 | 4501 | 4154 | 4154 | 7817 | 0.798 | 0.92 |
| 2016 | 1561 | 4689 | 3299 | 3299 | 6418 | 0.711 | 0.70 |
| 2017 | 496 | 2554 | 2237 | 2237 | 3639 | 1.052 | 0.88 |
| 2018 | 990 | 1783 | 1385 | 1385 | 2678 | 0.833 | 0.78 |
| 2019 | 2166 | 1289 |  |  |  |  |  |

Table 8.10.b. Cod in divisions 7e-k. Final XSA summary in relative value.

| Year | Relative Recruitment age 1 | Relative <br> tonnes | Landings <br> tonnes | Discards <br> tonnes | Relative <br> Ages |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0.87 | 1.05 | 5782 | 0 | 0.75 |
| 1972 | 0.169 | 0.97 | 4737 | 0 | 0.70 |
| 1973 | 0.51 | 0.90 | 4015 | 0 | 0.73 |
| 1974 | 0.162 | 0.87 | 2898 | 0 | 0.51 |
| 1975 | 1.10 | 0.79 | 3993 | 0 | 0.92 |
| 1976 | 0.36 | 0.76 | 4818 | 0 | 0.75 |
| 1977 | 0.52 | 0.92 | 3059 | 0 | 0.47 |
| 1978 | 0.50 | 1.01 | 3647 | 0 | 0.49 |
| 1979 | 1.21 | 1.03 | 4650 | 0 | 0.63 |
| 1980 | 2.2 | 1.08 | 7243 | 0 | 0.94 |
| 1981 | 0.94 | 1.17 | 10597 | 0 | 1.09 |
| 1982 | 0.38 | 1.41 | 8766 | 0 | 0.86 |
| 1983 | 1.26 | 1.36 | 9641 | 0 | 1.13 |
| 1984 | 1.22 | 1.00 | 6631 | 0 | 0.72 |
| 1985 | 1.07 | 1.37 | 8317 | 0 | 0.73 |
| 1986 | 0.91 | 1.43 | 10475 | 0 | 1.06 |
| 1987 | 4.6 | 1.19 | 10228 | 0 | 1.09 |
| 1988 | 2.2 | 1.73 | 17191 | 0 | 0.86 |
| 1989 | 0.66 | 2.7 | 19809 | 0 | 1.10 |
| 1990 | 0.73 | 2.00 | 12749 | 0 | 1.15 |
| 1991 | 2.1 | 1.13 | 9336 | 0 | 1.32 |
| 1992 | 2.1 | 0.95 | 9747 | 0 | 1.15 |
| 1993 | 0.67 | 1.28 | 10425 | 0 | 1.01 |
| 1994 | 2.5 | 1.50 | 10620 | 0 | 1.01 |
| 1995 | 1.76 | 1.36 | 11709 | 0 | 0.98 |
| 1996 | 1.35 | 1.66 | 12681 | 0 | 1.15 |
| 1997 | 1.82 | 1.47 | 12035 | 0 | 1.09 |
| 1998 | 0.91 | 1.31 | 11431 | 0 | 1.30 |


| Year | Relative Recruitment age 1 | Relative <br> tonnes | Landings <br> tonnes | Discards <br> tonnes | Relative <br> Ages |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 0.43 | 1.14 | 8594 | 0 | 1.20 |
| 2000 | 1.94 | 0.79 | 6536 | 0 | 1.05 |
| 2001 | 1.61 | 0.88 | 8308 | 0 | 1.04 |
| 2002 | 0.40 | 1.12 | 9236 | 0 | 1.17 |
| 2003 | 0.24 | 0.92 | 6420 | 0 | 1.28 |
| 2004 | 0.53 | 0.48 | 3672 | 0 | 1.31 |
| 2005 | 0.76 | 0.35 | 3062 | 0 | 1.35 |
| 2006 | 0.83 | 0.39 | 3776 | 0 | 1.14 |
| 2007 | 0.70 | 0.53 | 4830 | 0 | 1.14 |
| 2008 | 0.29 | 0.56 | 3961 | 0 | 1.04 |
| 2009 | 0.54 | 0.52 | 3292 | 0 | 1.06 |
| 2010 | 2.5 | 0.51 | 3229 | 0 | 0.78 |
| 2011 | 0.89 | 0.94 | 7261 | 696 | 0.68 |
| 2012 | 0.148 | 1.41 | 7692 | 952 | 0.94 |
| 2013 | 0.24 | 0.99 | 6290 | 530 | 1.33 |
| 2014 | 1.14 | 0.49 | 3879 | 741 | 1.20 |
| 2015 | 0.098 | 0.47 | 4154 | 565 | 1.08 |
| 2016 | 0.28 | 0.49 | 3299 | 220 | 0.96 |
| 2017 | 0.090 | 0.27 | 2237 | 117 | 1.42 |
| 2018 | 0.180 | 0.186 | 1385 | 180 | 1.12 |
| 2019 | 0.39** | 0.135 |  |  |  |

**25th quantile of the time-series (1971-2018).

Table 8.11. Cod divisions 7.e-k. Short-term forecast. Input table.

| Year | Age | N | M | Mat | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2019 | 1 | 2166 | 0.512 | 0.00 | 0.495 | 0.09345 | 0.908 |
| 2019 | 2 | 528 | 0.368 | 0.39 | 1.403 | 0.75854 | 2.098 |
| 2019 | 3 | 75 | 0.304 | 0.87 | 3.584 | 1.10657 | 4.231 |
| 2019 | 4 | 65 | 0.269 | 0.93 | 6.119 | 0.74357 | 7.006 |
| 2019 | 5 | 14 | 0.247 | 1.00 | 8.698 | 0.85300 | 9.031 |
| 2019 | 6 | 23 | 0.233 | 1.00 | 10.765 | 0.73038 | 10.138 |
| 2019 | 7 | 3 | 0.223 | 1.00 | 13.016 | 0.73038 | 11.649 |
| 2020 | 1 | 2166 | 0.512 | 0.00 | 0.495 | 0.09345 | 0.908 |
| 2020 | 2 | 1182 | 0.368 | 0.39 | 1.403 | 0.75854 | 2.098 |
| 2020 | 3 | 171 | 0.304 | 0.87 | 3.584 | 1.10657 | 4.231 |
| 2020 | 4 | 18 | 0.269 | 0.93 | 6.119 | 0.74357 | 7.006 |
| 2020 | 5 | 24 | 0.247 | 1.00 | 8.698 | 0.85300 | 9.031 |
| 2020 | 6 | 5 | 0.233 | 1.00 | 10.765 | 0.73038 | 10.138 |
| 2020 | 7 | 10 | 0.223 | 1.00 | 13.016 | 0.73038 | 11.649 |
| 2021 | 1 | 2166 | 0.512 | 0.00 | 0.495 | 0.09345 | 0.908 |
| 2021 | 2 | 1182 | 0.368 | 0.39 | 1.403 | 0.75854 | 2.098 |
| 2021 | 3 | 383 | 0.304 | 0.87 | 3.584 | 1.10657 | 4.231 |
| 2021 | 4 | 42 | 0.269 | 0.93 | 6.119 | 0.74357 | 7.006 |
| 2021 | 5 | 7 | 0.247 | 1.00 | 8.698 | 0.85300 | 9.031 |
| 2021 | 6 | 8 | 0.233 | 1.00 | 10.765 | 0.73038 | 10.138 |
| 2021 | 7 | 5 | 0.223 | 1.00 | 13.016 | 0.73038 | 11.649 |

Table 8.12. Cod divisions 7.e-k. Short-term forecast. Single option output table.

| Year | Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2019 | 1 | 0.09345 | 152 | 137.91 | 2166 | 1072 | 0 | 0 |
| 2019 | 2 | 0.75854 | 240 | 503.82 | 528 | 740 | 206 | 289 |
| 2019 | 3 | 1.10657 | 44 | 187.54 | 75 | 268 | 65 | 233 |
| 2019 | 4 | 0.74357 | 30 | 211.92 | 65 | 396 | 60 | 368 |
| 2019 | 5 | 0.85300 | 7 | 64.79 | 14 | 121 | 14 | 121 |
| 2019 | 6 | 0.73038 | 11 | 108.35 | 23 | 245 | 23 | 245 |
| 2019 | 7 | 0.73038 | 1 | 14.20 | 3 | 34 | 3 | 34 |
| 2019 | Total | 0.86500 | 485 | 1228.53 | 2874 | 2876 | 371 | 1290 |
| 2020 | 1 | 0.09345 | 152 | 137.91 | 2166 | 1072 | 0 | 0 |
| 2020 | 2 | 0.75854 | 538 | 1128.74 | 1182 | 1659 | 461 | 647 |
| 2020 | 3 | 1.10657 | 101 | 429.20 | 171 | 613 | 149 | 533 |
| 2020 | 4 | 0.74357 | 9 | 59.75 | 18 | 112 | 17 | 104 |
| 2020 | 5 | 0.85300 | 12 | 109.80 | 24 | 204 | 24 | 204 |
| 2020 | 6 | 0.73038 | 2 | 21.94 | 5 | 50 | 5 | 50 |
| 2020 | 7 | 0.73038 | 5 | 53.19 | 10 | 126 | 10 | 126 |
| 2020 | Total | 0.86500 | 819 | 1940.53 | 3576 | 3836 | 666 | 1664 |
| 2021 | 1 | 0.09345 | 152 | 137.91 | 2166 | 1072 | 0 | 0 |
| 2021 | 2 | 0.75854 | 538 | 1128.74 | 1182 | 1659 | 461 | 647 |
| 2021 | 3 | 1.10657 | 227 | 961.56 | 383 | 1374 | 333 | 1195 |
| 2021 | 4 | 0.74357 | 20 | 136.73 | 42 | 255 | 39 | 238 |
| 2021 | 5 | 0.85300 | 3 | 30.96 | 7 | 58 | 7 | 58 |
| 2021 | 6 | 0.73038 | 4 | 37.18 | 8 | 84 | 8 | 84 |
| 2021 | 7 | 0.73038 | 3 | 30.16 | 5 | 72 | 5 | 72 |
| 2021 | Total | 0.86500 | 947 | 2463.24 | 3793 | 4574 | 853 | 2294 |

Table 8.13. Cod divisions 7.e-k. Short-term forecast. Management options output.

| 2019 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| tsblnt | ssblnt | fmultint | fint |  | landInt |
| 2876 | 1290 |  |  | 0.8654181 | 1229 |
| tsb2020 | SSB_2020 | $F_{\text {mult }}$ | $F_{\text {bar }}$ | Landings | SSB_2021 |
| 3836 | 1664 | 0.0 | NA | 0 | 4447 |
| 3836 | 1664 | 0.1 | 0.08654 | 265 | 4144 |
| 3836 | 1664 | 0.2 | 0.17308 | 511 | 3864 |
| 3836 | 1664 | 0.3 | 0.25963 | 739 | 3607 |
| 3836 | 1664 | 0.4 | 0.34617 | 951 | 3371 |
| 3836 | 1664 | 0.5 | 0.43271 | 1147 | 3153 |
| 3836 | 1664 | 0.6 | 0.51925 | 1329 | 2952 |
| 3836 | 1664 | 0.7 | 0.60579 | 1499 | 2767 |
| 3836 | 1664 | 0.8 | 0.69233 | 1657 | 2596 |
| 3836 | 1664 | 0.9 | 0.77888 | 1804 | 2438 |
| 3836 | 1664 | 1.0 | 0.86542 | 1941 | 2293 |
| 3836 | 1664 | 1.1 | 0.95196 | 2068 | 2159 |
| 3836 | 1664 | 1.2 | 1.03850 | 2187 | 2035 |
| 3836 | 1664 | 1.3 | 1.12504 | 2299 | 1920 |
| 3836 | 1664 | 1.4 | 1.21159 | 2403 | 1814 |
| 3836 | 1664 | 1.5 | 1.29813 | 2500 | 1716 |
| 3836 | 1664 | 1.6 | 1.38467 | 2592 | 1625 |
| 3836 | 1664 | 1.7 | 1.47121 | 2677 | 1542 |
| 3836 | 1664 | 1.8 | 1.55775 | 2758 | 1464 |
| 3836 | 1664 | 1.9 | 1.64429 | 2833 | 1392 |
| 3836 | 1664 | 2.0 | 1.73084 | 2904 | 1325 |

Table 8.14. Catch option table.

| $F_{\text {mult }}$ | Catch20 | Land20 | Dis20 | FCatch20 | FLand20 | FDis20 | SSB21 | SSB.change20 | TAC.change19 | Advice.change20 | basis |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000000 | 0.0000 | 0 | 0.00000 | 0.00000 | NA | NA | 4447 | 167.20479 | -100.00000 | NA | f2019*0 |
| 0.1000000 | 285.0897 | 265 | 20.08965 | 0.08654 | 0.08654 | 0 | 4144 | 148.96811 | -82.26909 | Inf | f2019*0.1 |
| 0.2000000 | 549.7389 | 511 | 38.73891 | 0.17308 | 0.17308 | 0 | 3864 | 132.18769 | -65.84147 | Inf | f2019*0.2 |
| 0.3000000 | 795.0236 | 739 | 56.02359 | 0.25963 | 0.25963 | 0 | 3607 | 116.74187 | -50.61118 | Inf | f2019*0.3 |
| 0.4000000 | 1023.0953 | 951 | 72.09531 | 0.34617 | 0.34617 | 0 | 3371 | 102.51946 | -36.48128 | Inf | f2019*0.4 |
| 0.5000000 | 1233.9541 | 1147 | 86.95407 | 0.43271 | 0.43271 | 0 | 3153 | 89.41885 | -23.36308 | Inf | f2019*0.5 |
| 0.6000000 | 1429.7515 | 1329 | 100.75149 | 0.51925 | 0.51925 | 0 | 2952 | 77.34716 | -11.17539 | Inf | f2019*0.6 |
| 0.7000000 | 1612.6392 | 1499 | 113.63919 | 0.60579 | 0.60579 | 0 | 2767 | 66.21943 | 0.15614 | Inf | f2019*0.7 |
| 0.8000000 | 1782.6172 | 1657 | 125.61717 | 0.69233 | 0.69233 | 0 | 2596 | 55.95797 | 10.69959 | Inf | f2019*0.8 |
| 0.9000000 | 1940.7612 | 1804 | 136.76124 | 0.77888 | 0.77888 | 0 | 2438 | 46.49170 | 20.51736 | Inf | f2019*0.9 |
| 1.0000000 | 2088.1472 | 1941 | 147.14721 | 0.86542 | 0.86542 | 0 | 2293 | 37.75553 | 29.66664 | Inf | f2019*1 |
| 1.1000000 | 2224.7751 | 2068 | 156.77508 | 0.95196 | 0.95196 | 0 | 2159 | 29.68990 | 38.19985 | Inf | f2019*1.1 |
| 1.2000000 | 2352.7965 | 2187 | 165.79647 | 1.03850 | 1.03850 | 0 | 2035 | 22.24028 | 46.16508 | Inf | f2019*1.2 |
| 1.3000000 | 2473.2872 | 2299 | 174.28719 | 1.12504 | 1.12504 | 0 | 1920 | 15.35669 | 53.60647 | Inf | f2019*1.3 |
| 1.4000000 | 2585.1714 | 2403 | 182.17143 | 1.21159 | 1.21159 | 0 | 1814 | 8.99336 | 60.56449 | Inf | f2019*1.4 |
| 1.5000000 | 2689.5250 | 2500 | 189.52500 | 1.29813 | 1.29813 | 0 | 1716 | 3.10833 | 67.07633 | Inf | f2019*1.5 |
| 1.6000000 | 2788.4995 | 2592 | 196.49952 | 1.38467 | 1.38467 | 0 | 1625 | -2.33684 | 73.17614 | Inf | f2019*1.6 |
| 1.7000000 | 2879.9434 | 2677 | 202.94337 | 1.47121 | 1.47121 | 0 | 1542 | -7.37742 | 78.89528 | Inf | f2019*1.7 |
| 1.8000000 | 2967.0840 | 2758 | 209.08398 | 1.55775 | 1.55775 | 0 | 1464 | -12.04576 | 84.26257 | Inf | f2019*1.8 |
| 1.9000000 | 3047.7697 | 2833 | 214.76973 | 1.64429 | 1.64429 | 0 | 1392 | -16.37152 | 89.30452 | Inf | f2019*1.9 |
| 2.0000000 | 3124.1522 | 2904 | 220.15224 | 1.73084 | 1.73084 | 0 | 1325 | -20.38193 | 94.04550 | Inf | f2019*2 |
| 0.0653326 | 189.3426 | 176 | 13.34256 | 0.05654 | 0.05654 | 0 | 4246 | 155.11809 | -88.26193 | Inf | msyap- |
| 0.4044288 | 1032.7776 | 960 | 72.77760 | 0.35000 | 0.35000 | 0 | 3361 | 101.91623 | -35.87952 | Inf | msy |
| 0.1026671 | 292.6203 | 272 | 20.62032 | 0.08885 | 0.08885 | 0 | 4136 | 148.50133 | -81.81366 | Inf | msy_up- |
| 0.0429388 | 124.7940 | 116 | 8.79396 | 0.03716 | 0.03716 | 0 | 4314 | 159.18808 | -92.22027 | Inf | msy_lower |
| 0.7696627 | 1732.0541 | 1610 | 122.05410 | 0.66608 | 0.66608 | 0 | 2646 | 58.98331 | 7.58100 | Inf | TACstable |
| 0.9341958 | 1991.3243 | 1851 | 140.32431 | 0.80847 | 0.80847 | 0 | 2387 | 43.42560 | 23.71815 | Inf | TACplus15 |
| 0.6224737 | 1471.7081 | 1368 | 103.70808 | 0.53870 | 0.53870 | 0 | 2909 | 74.76644 | -8.55615 | Inf | TACmi- |
| 0.6701963 | 1559.9245 | 1450 | 109.92450 | 0.58000 | 0.58000 | 0 | 2820 | 69.44191 | -3.13567 | Inf | pa |
| 0.9244087 | 1977.3388 | 1838 | 139.33878 | 0.80000 | 0.80000 | 0 | 2402 | 44.29432 | 22.81017 | Inf | lim |
| 1.0000022 | 2088.1472 | 1941 | 147.14721 | 0.86542 | 0.86542 | 0 | 2293 | 37.75553 | 29.66664 | Inf | fint |
| 1.0000022 | 2088.1472 | 1941 | 147.14721 | 0.86542 | 0.86542 | 0 | 2293 | 37.75553 | 29.66664 | Inf | Btrigger |
| 1.0000022 | 2088.1472 | 1941 | 147.14721 | 0.86542 | 0.86542 | 0 | 2293 | 37.75553 | 29.66664 | Inf | Bli |



Figure 8.1. Cod in divisions 7e-k. Historical landings (in Tonnes) by country.


Figure 8.2. Cod in divisions 7e-k. Catches volume in Tonnes (i.e. landings plus discards) by area, season and country in 2018.


Figure 8.3.a. Cod in divisions 7e-k. Imported French 2018 landings and discards length distribution.


Figure 8.3.b. Cod in divisions 7e-k. Imported Irish 2018 landings and discards length distribution.


Figure 8.3.c. Cod in divisions 7e-k. Imported Belgium 2018 landings and discards length distribution.


Figure 8.3.d. Cod in divisions 7e-k. Raised United Kingdom 2018 landings length distribution - Sampled strata only.


Figure 8.4. Cod in divisions 7e-k. Raised age distribution of the catches (landings and discards) in 2018.


Figure 8.5a. Cod in divisions 7e-k. Time-series of (a) lpue and (b) fishing effort for the French fleets. Units: Lpue in kg/day and fishing effort in days fished.


Figure 8.5b. Cod in divisions 7e-k. Time-series of (a) Standardized Ipue and (b) fishing effort for the Irish fleets. Units: Ipue in $\mathrm{kg} /$ day fished and Effort in 000s hours fished.


Figure 8.5c. Cod in divisions 7e-k. Time-series of (a) lpue and (b) fishing effort for the UK fleets. Units: Ipue in kg/day and fishing effort in days fished.


Figure 8.6. Cod in divisions 7e-k. Tuning indices used in the assessment. Commercial tuning fleet corresponds to French OTDEF Q2+3+4 where number-at-age are plotted. The survey index is a combined index based on both French IR-GFS and FR-Evhoe Q4 data where mean number-at-age per hour and grid cell are plotted.


Figure 8.7. Cod in divisions 7e-k. Final assessment. Residuals (Left panel: French OTDEF demersal tuning fleet; Right Panel: Combined survey indices).


Figure 8.8. Cod in divisions 7e-k. Final XSA outputs. Fishing mortality. $\mathrm{F}_{\text {bar }}=$ Thick black line. Age $\mathbf{0}$ are not included in the assessment.



Figure 8.9. Cod in divisions 7e-k. Final XSA outputs. (a) Catch and (b) stock number-at-age. Age 0 are not included in the assessment.

Figure 8.10. Cod in divisions 7e-k. Final XSA outputs. Summary plots. Recruitment, F, and SSB values.


Figure 8.11a. Cod in divisions 7e-k. Final XSA. Retrospective plots.

## Cod in VIlek



F in legend for year shown by vertical dotted line

8.11b. Cod in divisions 7e-k. Final XSA. Comparison between runs (runs with the two tuning indices, with only the survey index and with only the commercial tuning index).

## Landings yield 2020

SSB 2021


Figure 8.12. Cod in divisions 7e-k. Forecast yield in 2020 and SSB in 2021.

## 7 Haddock (Melanogrammus aeglefinus) in Division 6.b (Rockall)

## Type of assessment in 2019: Update assessment taking into account the recommendations of benchmark 2019

The current assessment is an update of last year's assessment assessment taking into account the recommendations of benchmark 2019. The same approach has been used in the annual assessment since 2005 when on the recommendation of RGNSDS, adopted a new assessment approach, which allows modelling of the total catch (including discards) when no on-board observations were available (for details see the Stock Annex).

## ICES advice applicable to 2019

ICES advice applicable to 2019 can be found here:
http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2018/2018/had-rock.pdf

### 7.1 General

## Stock description and management units

The haddock stock at Rockall is an entirely separate stock from that inhabiting the continental shelf of the British Isles. Since 2004, the EU TAC for haddock in $6 . \mathrm{b}$ has been included with divisions 12 and 14 . For details of the earlier management units see the Stock Annex.

## Management applicable to 2018 and 2019

The EU TAC for 6.b, 12 and 14 was set at 5163 t in 2018 (a $10 \%$ increasing compared to TAC for 2017).

| Species:Haddock <br> Melanogrammus aeglefinus Zone: Union and international waters of $6 \mathrm{~b}, 12$ and 14 <br> (HAD/6B1214) <br> Belgium 12  <br> Germany 40  <br> France 546  <br> Ireland 429  <br> United Kingdom 4136  <br> Union 5163 Analytical TAC <br> TAC 5163 Article 7(2) of this Regulation applies |
| :--- | ---: | :--- | :--- |

The EU TAC for 6.b, 12 and 14 was set at 10469 t in 2019 (a $202 \%$ increasing compared to TAC for 2018).

| Species:Haddock <br> Melanogrammus aeglefinus Zone: Union and international waters of $6 \mathrm{~b}, 12$ and 14 <br> (HAD/6B1214) <br> Belgium 23  <br> Germany 28  <br> France 1155  <br> Ireland 824  <br> United Kingdom 8439  <br> Union 10469 Analytical TAC <br> TAC 10469 Article 7(2) of this Regulation applies <br>    |
| :--- | :--- | :--- | :--- |

The ICES advice, agreed TAC for EU waters, and WG estimates of landings during 2002-2018 is summarised below. All values are in thousand tonnes.

| YEAR | Predicted catch corresp. to advice | Predicted landings corresp. to advice\# | BASIS | AGREED <br> TAC a | WG LANDINGS |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | $<1300$ |  | Reduce F below 0.2 |  | 3.336 |
| 2003 | - |  | Lowest possible F |  | 6.242 |
| 2004 | - |  | Lowest possible $\mathrm{F}^{\text {b }}$ | 0.702 | 6.445 |
| 2005 | - |  | Lowest possible $\mathrm{F}^{\text {b }}$ | 0.702 | 5.179 |
| 2006 | - |  | Lowest possible $\mathrm{F}^{\text {b }}$ | 0.597 | 2.765 |
| 2007 | $<7100$ |  | Reduce F below $\mathrm{F}_{P A}{ }^{\text {b }}$ | 4.615 | 3.349 |
| 2008 | <10 640 |  | Keep $F$ below $\mathrm{F}_{P A}{ }^{\text {b }}$ | 6.916 | 4.221 |
| 2009 |  | <4300 | No long-term gains in increas- | 5.879 | 3.445 |
| 2010 |  | <3300 | Little gain on the long-term | 4.997 | 3.405 |
| 2011 |  | $<2700$ | Reduction in F is needed to | 3.748 | 1.903 |
| 2012 |  | <3300 | MSY approach | 3.300 | 0.710 |
| 2013 | 0 | 0 | No directed fisheries, mini- | 0.990 | 0.826 |
| 2014 | <1620 | <0980 | MSY approch | 1.210 | 1.675 |
| 2015 | $<4310$ | <2930 | MSY approch | 2.580 | 2.445 |
| 2016 | <3932 | <3225 | MSY approach | 3.225 | 2.585 |
| 2017 | $\leq 4690$ | $\leq 4130$ | MSY approach | 4.690 | 4.610 |
| 2018 | $\leq 5163$ |  | MSY approach | 5.163 | 3.868 |
| 2019 | $\leq 10469$ |  | MSY approach | 10.469 |  |

${ }^{a}$ Before 2014 TAC was set for Divisions 6 .a and $6 . b$ (plus 5 b1, 12 and 14) combined with restrictions on quantity that can be taken in 5.b and 6.a. The quantity shown here is the total area TAC minus the maximum amount which is allowed to be taken from 5.b and 6.a. In 2004, the EU TAC for Division 6 was split and the $6 . \mathrm{b}$ TAC for haddock was included with XII and XIV. This value is the TAC for 6.b, 12 and 14.
${ }^{\text {b }}$ Single-stock boundary and the exploitation of this stock should be conducted in the context of mixed fisheries, protecting stocks outside safe biological limits.
\# Wanted catch.

The minimum landing size of haddock taken by EU vessels at Rockall is 30 cm . There is no minimum landing size for haddock taken by non-EU vessels in international waters.

In order to protect the pre-recruit stock, the International Waters component of the statistical rectangle 42D5 has been closed for fishing since 2001 and its EU component since 2002 (see the

Stock Annex). The protected area (the whole rectangle) is referred to as Rockall Haddock Box. In order to protect cold-water corals, three further areas (North West Rockall, Logachev Mounds and West Rockall Mounds) were closed since January 2007 (see the Stock Annex). A new area to protect cold-water corals (Empress of British Banks) was established by the NEAFC in 2007 and 2012.

Since 2009 in NEAFC regulatory area, including international waters of Rockall, was established a ban on discards.

## Fishery in 2018

## Russian fishery in 2018

Directed fishing of haddock in Rockall by Russian vessels was not conducted in 2018.

## Scottish fishery in 2018

Total Scottish haddock landings in $6 . b$ in 2018 were estimated to be 3418 t (Table 4.3.1). Other important target species included anglerfish (Lophius spp.), ling, saithe and megrim. Scottish effort presented in Table 4.3.2 and 4.3.3.

## Irish fishery in 2018

Irish effort in Rockall declined in 2009-2015 and increased again in 2016-2018 (Table 4.3 .2 and 4.3.3).

Landings totalling 433 thaddock were reported from Irish otter trawlers in 2018 (decreased from 500 t in 2017; Table 4.3.1). Irish vessels used single otter trawls with a mesh size ranging from 100 to 120 mm together with a square mesh panel.

## Norwegian fishery in 2018

In 2014-2016 Norwegian landings of haddock at Rockall 41-66 t and 26 t in 2017 were reported. Total Norwegian landings of 16 t of haddock at Rockall were reported in 2018. Norwegian demersal fleet fishing on the Rockall Bank consisted mainly of longliners targeting mainly ling and tusk.

### 7.2 Data

## Landings

Nominal landings as reported to ICES are given in Table 4.3.1, along with Working Group estimates of total estimated landings. Revisions to official catch statistics for previous years are also shown in Table 4.3.1.

Anecdotal evidence suggests that misreporting of haddock from Rockall has occurred historically (which may have led to discrepancies in assessment), but a quantitative estimation of the degree of misreporting is not possible.

International age composition and mean weight-at-age in the landings were compiled according to the methods described in the Stock Annex.

## BMS landings

In 2016 BMS (Below Minimum Size) landings which are subject EU landings obligation were only 0.4 t . In the assessment BMS landings were include in total landings. In 2017-2018 BMS landings were not reflected in the catch statistic.

## Discards

Historically, the discard rate was as high as $12-87 \%$ by numbers according to the results of discards trips (see the Stock Annex). The methods used to reconstruct the historical time-series of discards is described in the Stock Annex.

The discards for 2010-2016 in the 2017 assessment were estimated from sampling aboard Scottish and Irish vessels collected in 2010-2016 (Table 4.3.4-4.3.6). On Russian vessels, the whole catch of haddock is kept on board and therefore, total catch is equivalent to landings and there is no need to calculate discards. In 2015, the discard rate was estimate at $38 \%$ and $52 \%$ by numbers on Scottish and Irish observer trips. In 2016, the level of discards has not changed significantly and was estimate at $11 \%$ and $56 \%$ by numbers on Scottish and Irish observer trips.

In 2017, the discard rate was also high and was estimated at $17 \%$ by numbers on Scottish and $39 \%$ on Irish observer trips. In 2018, discard rate was estimated at $32 \%$ by numbers on Scottish and $32 \%$ on Irish fleet (Table 4.3.4-4.3.7).

## Biological

There was no change in biological parameters compared to the 2018 assessment (see the Stock Annex).

## Surveys

There is only one abundance index available for this stock the Scottish Rock-IBTS-Q3 survey (Figure 4.3.1-3). The survey is coordinated by IBTS and described further in the IBTS reports and Stock Annex.

The area which was covered by the survey was not stable and moreover the survey coverage has been extended in recent years (Figure 4.3.1). The 2019 indices were obtained from the standard survey area as last year's. During the benchmark 2019, the number of different runs were conducted to explore the sensitivity of the assessment results and diagnostics to different survey indices. The correction to the survey data has little impact on the results of the stock assessment (Figure 3). In terms of diagnostics (Figure 4 and 5), the fit to the 2015 survey data at age 5 shows significant improvement in Run S1 compared to the baseline: in the baseline model run there is a large positive residual which is no longer apparent. The benchmark agreed that based on the model diagnostics, the 'standard' survey index with corrected ALK for ages 4-5 in 2015 should continue to be used in the assessment (Table 4.3.8).

Additional abundance and biomass estimates are calculated by the swept area method using three types of stratification of the survey area:

1. by geographic strata of 15 ' latitude wide and 15 ' longitude long (Figure 4.3.4);
2. by five bathymetric strata depending on depth: <150 m, 150-175 m, 176-200 m, 201-225 m and $>225 \mathrm{~m}$ (Figure 4.3.5);
3. the whole survey area is taken for one strata without substratification (Figure 4.3.6).

All three methods show similar patterns (Figures 4.3.4-6).
In 2011, the gear was changed on the Scottish survey and an analysis showed that there was no detectable difference between the older and new survey on haddock indices in neighbouring areas (IBTSWG 2012).

The Russian trawl acoustic survey conducted in 2005 provided information on the size and biomass of the haddock stock both in the EU zone and in international waters. The acoustic survey yielded a biomass estimate of 60000 t and an abundance estimate of 225.9 million (for the details see the Stock Annex). No such survey has been conducted in subsequent years.

## Commercial Effort, Lpue and Cpue

Commercial effort series are available for Scottish trawlers, light trawlers, seiners, Irish otter trawlers and Russian trawlers fishing in Division 6.b. The effort data for these fleets are shown in Figure 4.3.7 and Table 4.3.2-4.3.3. Effort data in hours from the Scottish fleets is discontinued after 2008 and provided in KWDays after 2003 (Table 4.3.3. Effort by the Scottish and Irish fleets has been relatively stable at a low level in the last three years.

Commercial Lpue for the Irish and Scottish fleets and cpue for the Russian fleet are shown in Figures 4.3.8-4.3.9. The WG decided that the commercial cpue and lpue data, which do not include discards and have not been corrected for changes in fishing power despite known changes in vessel size, engine power, fish-finding technology and net design, were unsuitable for catch-at-age tuning.

### 7.3 Description of stock assessment approach

Model used:
The assessment is based on catch-at-age data and one survey index (Scottish Rock-IBTS-Q3) and conducted using the XSA method.

Software used:

In 2005-2018 same software was used as in the last year's assessment (XSA from Lowestoft suite of VPA programs). In 2019 taking into account the recommendations of benchmark 2019 FLXSA was used

Model Options chosen:
Settings for the final XSA assessment did not change compared to the previous assessment (see the Stock Annex) and were as follows:

Assessment model: XSA
Tuning indices: one survey index (Scottish Rock-IBTS-Q3)
Time-series weights: none
Catchability dependent for ages $<4$
Regression type: C
Minimum number of points used for regression: 10
Q plateau: 5
Shrinkage stand. error: 1.0
Shrinkage age, year: 4 years, 3 ages
Minimum stand. error: 0.3
Plus group: 7+
Fbar: 2-5

Input data types and characteristics:
There were no changes in data types and characteristics compared to the previous assessment:
Year range: 1991-2018
Age range: 1-7+
For tuning data the following year and age ranges were used:
Year range: 1991-2018
Age range: 1-6

## Data screening

Figures 4.3.10 and 4.3.11 as well as Tables 4.3 .9 show landings, discards and total catch by number and weight. Landings, discards and total catch-at-ageby number are shown in Tables 4.3.104.3.12.

In 2016, BMS (Below Minimum Size) landings which are subject EU landings obligation were only 0.4 t . In assessment BMS landings were include in total landings. Because BMS landings were low, this did not lead to a decrease in the average weight of landings.
For 2012 onwards, the catch-at-age data were estimated in InterCatch. The main fleets (UK(Sco) OTB_DEF_>=120 and Irish OTB_DEF_100-119) are typically sampled for both landings and discards. Discard rate allocation to other unsampled fleets consisted of:

- Manually matching annual discards to available quarterly landings by country/fleet (where necessary).
- Using a weighted average discard rate for all unsampled fleets (weighted by CATON) with the exception of the Norwegian longline fleet and the Russian fleet for which discards are both assumed to be zero..

Landings age compositions were allocated to unsampled fleets using a weighted average of all sampled fleets (excluding the Russian fleet which is likely to be less applicable given they do not discard). The weighting algorithm used is ‘Mean weight weighted by numbers-at-age or length'. Discards age compositions were allocated in a similar manner.

In 2019, a benchmark assessment was conducted on this stock. During the benchmark meeting, the discards data for 2010 was also revised. Before 2019, the discards for 2010 were calculated on average discards proportion in total catch but that proportion was applied on the landings. However it is not correct because discards proportion is relative the total catch. That method not correct reflected discards because not take into account that the haddock at age 1 and age 2 mostly discarded and not reflect in landings. As result, landings in 2010 were same as in 2009 but discard in 2010 decline to much and the discards proportion in 2010 declined compare to previous years. In benchmarked assessment 2019, the discards for 2010 were calculated compared to 2009 with same proportion-at-age as proportion between landings in 2009 and 2010.
The resulting age compositions and mean weights-at-age show only minor differences to those compiled at previous assessment WG meetings (Figure 4.3.12).
Mean weights-at-age in total catch, landings, discards and stock are shown in Tables 4.3.13-4.3.16 and Figures 4.3.13-4.3.16.
Historically, stock weights-at-age have been assumed to be equal to the raw catch weights. In recent years, the number of sampled trips for both landings and discards has been verylow. This lead to higher variability in the mean weight-at-age estimates. For this reason, recent years the smoothed catch weights-at-age was applied by the WGCSE.

To mitigate against variability in the mean weight-at-age since 2019 mean weights in the stock are assumed as five-year means taking into account the recommendations of benchmark 2019.

The mean weights-at-age in the total catch (including discards) and in the stock are shown in Figure 4.3.14.

There were small landings of haddock aged 1 in 2010-2012 and very few aged 2 to 6 compared to historical values. Haddock aged 7 dominated landings. However, in 2013 landings and discards of haddock aged 1 significant increased. Discarded fish are, primarily, haddock aged 1-2 (see Tables 4.3.1 and 4.3.2 in the Stock Annex). Figures of log catches by age show that these values are much less variable when discards are included (Figures 4.3.15-4.3.20). Data on catches, landings and discards-at-age are given in Tables 4.3.10-4.3.13.

The Scottish Rock-IBTS-Q3 was the only survey index available to the working group. Plots of $\log$ cpue by age, year and year class are shown in Figures 4.3.21-4.3.24.

Comparative scatter plots of log index at-age are shown in Figure 4.3.24. The survey shows relatively good internal consistency in tracking year-class strength through time.

## Final update assessment

## Final run

Settings for the final XSA assessment are shown in Section C of the Stock Annex. There have been no changes to assessment settings since 2013.

The diagnostics file of the final XSA run is given in Table 4.3.17 and Figure 4.3.25. Adjusted survey cpue against XSA population estimates are shown in Figures 4.3.26, 4.3.26. The analysis of residuals and retrospective analysis (Figures 4.3.25,4.3.28) show that applying the chosen parameters for XSA (as in the Stock Annex) improves the residual patterns compared to other exploratory settings. However, the same trends are still apparent in the log catchability residuals. The results of the retrospective analysis conducted by the Working Group in 2002,2003 and 2005 indicated that using shrinkage values of more than 0.5 improved the retrospective curves and showed convergence. In this year's analysis, only 28 years data were available for the retrospective analysis, but a good year-to-year consistency was obtained. The final XSA results are given in Tables 4.3.18-4.3.20.

Summary plots from the final XSA assessment are shown in Figure 4.3.29.

## Comparison with previous assessments

The estimates from this year's assessment are reasonably consistent with the assessments carried out in previous years (Figure 4.3.30). In 2019, a benchmark assessment was conducted on this stock. The new assessment resulted in the SSB assessed for 2017 to be revised upwards by $10 \%$ and recruitment was increased by $25 \%$ compared to last year's assessment. In addition fishing mortality for 2010 been revised upwards in this year's assessment. It is result of revision of discard data for 2010. As result, total catch was estimated close to 2009 same as landings. The decline of SSB in 2010 with the same the catch gave increase F.

## State of the stock

The stock summary relative to reference points is plotted in Figure 4.3.29.
The spawning-stock biomass (SSB) has increased from the lowest observed in 2014 and is estimated to be above MSY $\mathrm{B}_{\text {trigger }}$ in 2016. Fishing mortality $(\mathrm{F})$ has declined over time but has been above Fmsy in 2018. Recruitment during 2008-2012 is estimated to be extremely weak. Recruitment has improved in 2013-2014 and decreased again in 2015-2016 and is still lower than the
values estimated at the beginning of the time-series. The 2017 year class estimated by results of survey was on high level. In 2018, recruitment has decreased again.

## Statistical catch-at-age analysis (SCAA)

For Statistical catch-at-age analysis, StatCam model was used (J. Brodziak, 2005). VPA and SCAA used identical survey and catch data. For StatCam runs two scenarios were used: First scenario, non-parametric model; second, parametric model.
The StatCam model shows good conformity between observed and predicted survey index and catch biomass (Figure 4.3.31).
StatCam summary plots are shown in Figure 4.3.32.
Both Statistical catch-at-age analysis and VPA results show a similar tendency for the SSB dynamics. However, the assessment of the stock size depends on the choice of the model.

## Results of stock assessment by SAM model (state-space assessment model)

For both runs of SAM and the VPA was used same input data. The summary plots for SAM assessment are shown in Figure 4.3.33.

The SAM assessment and VPA results show a similar tendency in state stock. However, recent years the SSB assessed by the SAMincreased slowly compare to the VPA assessment.

## The comparison of stock assessment results produced by different models

All that models results show a similar tendency for the SSB dynamics. However, there are variation in the interannual dynamics of the stock assessed by different models.

The SSB and TSB plots from the XSA, SCAA and SAM assessment are compared in Figure 4.3.34.

### 7.4 Short-term projections

## Estimating year-class abundance

In 2007-2011, the abundance of age 0 individuals in the survey index were estimated to be extremely weak. In 2012, the observed large number 0-group. Year classes 2013 and 2014 were below average but above levels 2008-2012 (Figure 4.3.35). No significant relationship between spawning biomass and the recruitment was found. Poor year classes may be related to environmental factors including rising seawater temperatures in Rockall Bank, a reduction in zooplankton abundance (ephausiids and Calanus finmarhicus) and the negative impact of predation on eggs and larvae and food competition from the grey gurnard. The 2012 year class, was overestimated by the survey assessment of 0-group. It resulted in an increase of uncertainty in the assessment because more than $70 \%$ of 0 -group fish were caught during a single haul (Figure 4.3.2). In 2007-2016 the recruitment (age 1) assessed by VPA was below average for full time-series 1991-2016 (Table 4.3.20).

In 2016 and 2017, a strong 0-group was observed. But a considerable number of 0-group fish were caught during a single haul (Figure 4.3.2). This increases the uncertainty of forecasting recruitment as in 2012.

VPA abundance for age 1 has been highly correlated with age 0 indices for 1993-2015 ( $\mathrm{r}^{2}=0.75$ ) but in 2016-2017 this correlation declined (Figure 4.3.36). The recruitment (age 1) in 2013-2017 was therefore estimated using RCT3 regression (Shepherd, 1997) relating survey indices to stock abundance. The recruitment in 2019 was estimated at 24444 thousand.

For forecasting recruitment (age 1) in 2020 and thereafter, the WG recommended the same procedure as last year using the 25th percentile over the whole time-series.

Many definitions of how to compute the percentile maybe found in the literature. TheWG chose the simple rounding of the result to the nearest integer and taking the value that corresponded to that rank of percentile. The rank of percentile was determined by the following equation:

$$
n=\frac{P}{10} * N+\frac{1}{2}
$$

$P$ being the percentile value (here $\mathrm{P}=28$ ), and N the length of the time-series (here $\mathrm{N}=27$ ). The rank of 25 th percentile for the recruitment is then 8 . The 7 th lowest value of the time-series corresponds to a value of 14170 thousands in 2004.

The input data for the short-term forecast can be found in Table 4.3.21.

## Catch Constraint

A catch constraint is used for 2019. The assumed catch in 2019 of $9763 t$ is estimated based on UK (8439t) and Irish (824t) quotas and an estimated Russian catch 500 t . Recent UK and Irish quota up take has been high and the Russian fishery has already taken place in 2019 so the catch constraint forecast, as last year, is considered to be the best approach by the WG.

Results of forecast are shown in Tables 4.3.21-4.3.23.

## Mean Weights and F pattern

In recent years, the number of sampled trips for both landings and discards has been very low. This leads to higher variability in catch and survey estimates of those year classes, increasing the uncertainty in F. Since 2015, to mitigate against this in the forecast five-year averages for weight were used in the catch options. Average of the three last years of exploitation patterns and weight-at-age in the stock were used.

## Partitioning of catch into discards and landings

An important uncertainty in the assessment and forecast concerns the estimates of discards. The number of sampled discard trips in the last years has been very low. Since the discard ratio at age varies considerably from year to year a ten-year average discard proportion (2009-2018) was used for forecasting discards in the short term (Tables 4.3.7-4.3.10 and Figure 4.3.37).

## STF results

Results obtained from the forecast (including discards) are given in Tables 4.3.21-4.3.23.
Stock numbers of recruits and their source for recent year classes used in the predictions and the relative (\%) contributions to landings and SSB (by weight) of these year classes are shown in Table 4.3.24.

### 7.5 MSY evaluations and Biological reference points

ICES carried out an evaluation of MSY and PA reference points for this stock in 2019 WKROCKMSE (ICES, 2019). The results are summarized below:

| Frame- <br> work | Reference <br> point | Value | Technical basis | Source |
| :--- | :--- | :--- | :--- | :--- |
| MSY ap- <br> proach | MSY Btrig- <br> ger | 3712 tonnes | Bpa | ICES (2019) |
|  | FMSY | 0.168 | Segmented regression with Bloss, the lowest <br> observed spawning-stock biomass (EqSim). | ICES (2019) |
| Precau- <br> tionary <br> approach | Blim | 2474 tonnes | Blim = Bloss, the lowest observed spawning- <br> stock estimated in previous assessments. | ICES (2019) |


| Bpa | 3712 tonnes | Bpa = Blim $\times 1.4$. This is considered to be the <br> minimum SSB required to obtain a high proba- <br> bility (95\%) of maintaining SSB above Blim | ICES (2019) |
| :--- | :---: | :--- | :--- | :--- |
| Flim | 1.06 | Based on a 50\% probability of being above Blim <br> in a stochastic simulation with a segmented re- <br> gression using breakpoint at Blim. | ICES (2019) |
| Manage- <br> ment <br> plan SSBmgt 3712 tonnes Bpa | ICES (2019) |  |  |


|  | Fmgt | 0.168 | Based on harvest control rule evaluations. | ICES (2019) |
| :--- | :--- | :---: | :--- | :--- |
| Manage- <br> ment <br> plan* | MAP <br> MSY Btrig- <br> ger | 3712 tonnes | MSY Btrigger |  |


| MAP Blim | 2474 tonnes | Blim |  |
| :--- | :---: | :--- | :--- |
| MAP FMSY | 0.168 | FMSY | ICES (2019) |
| MAP range <br> Flower | 0.105 | Consistent with ranges provided by ICES <br> (2016a), resulting in no more than 5\% reduction <br> in long-term yield compared with MSY. |  |


| MAP range <br> Fupper | 0.27 | Consistent with ranges provided by ICES <br> (2016a), resulting in no more than 5\% reduction <br> in long-term yield compared with MSY. | ICES (2019) |
| :--- | :--- | :--- | :--- |

### 7.6 Management plans

In September2011 and 2012 in accordance with the conclusions of the 2010-2011 Annual Meeting of the NEAFC, a delegation from the RF and EU considered the management plan. In the light of the ICES comments, the necessary adjustments required to the draft plan were considered. The revised proposal for a harvest control component of a long-term management plan for haddock at Rockall was forwarded to NEAFC at the opportunity for approval at the 2012 Annual Meeting. ICES is requested toevaluate the EU-Russia proposal for theharvestcontrol component of the management plan for Rockall haddock and to evaluate the proposals on the protection of juvenile Rockall haddock. According the management plan the measure shall be put in place to ensure that total catch does not exceed the established TAC including measures to record and
minimise discards. It is the consideration of 2004 Expert Group the basic measure to reduce discards should be effort regulation along with the biological reasonable the minimum landings size.

ICES evaluated a new HCR proposal RF and EU for the Rockall haddock stock in August 2013 (ICES, 2013) and found that a maximum F of 0.2 was required in the HCR to ensure consistency with the precautionary approach, under the low recruitment conditions observed since 2004.

The management plan additionally indicates that measures should be put in place to ensure that total catch does not exceed the established TAC, including measures to record and minimize discards. After the introduction of these measures, the human consumption TAC method currently used by ICES (advice based on landings) should not be applied.

In 2017, NEAFC again requested ICES to evaluate the following proposal for the harvest control component of a long-term management plan for Rockall haddock and in particular to consider whether the plan is consistent with the precautionary approach and will provide for the sustainable harvesting of the stock. If the plan fails to be precautionary ICES will also be asked to suggest possible options to bring the plan aligned with the precautionary approach. In 2019, ICES evaluated proposed NEAFC plan.

By NEAFC opinion the measures to reduce discards for whole area distribution of stock need to develop and to implement in practice, while also reducing the TAC to take into account any discarding that is still taking place for realization of management plan. In NEAFC regulatory area (RA) established a ban on discards. The remainder of the management plan for this species is considered to be suitable and has been agreed by the Contracting Parties (NEAFC, 2015).

### 7.7 Uncertainties and bias in assessment and forecast

The WG considers that the long-term trends in the XSA assessment and survey biomass estimates/indices are indicative of the general stock trends. The assessment has become increasingly uncertain in recent years as catch and sampling levels have declined to low levels. To mitigate against variability in the mean weight-at-age in the stock are assumed as five-year means. The three-year averages of exploitation patterns and five-year average catch weights and ten-year mean discard proportions were used in the catch options for forecast.

### 7.8 Recommendation for next benchmark

In recent years WGCSE have highlighted an increasing number of issues to be addressed when this stock is benchmarked.

1. The WG considers that a longer series of more accurate landings, discards (for non-Russian fleets) and survey data will be necessary to overcome these deficiencies.
2. There are concerns over the accuracy of landings statistics from Rockall in earlier years. There was no analysis of which method is better to use when in terms poor information by result discards trips: the method of estimating discards from survey data or the results poor discards, especially in 2010 where an average rate had to be used since the survey could not take place.
3. In 1999 and 2011, the gear and tow duration were changed on the Scottish survey. Analysis of the impact of this on the stock assessment is needed.
4. The XSA assessment shows trends in catchability, even if reduced by weak shrinkage. Diagnostics give quite large standard errors on survivors' estimates (0.3-0.4) and there are often quite different values given by Scottish Rock-IBTS-Q3, F-shrinkage and P-
shrinkage. During benchmark 2019, progress has been made but further efforts in this direction are needed.
5. The survey covers only part of the currently known distribution area of haddock that raises uncertainty in the assessment.
6. The indices obtained from the standard survey area must be used for the next assessment on account of the heterogeneity in the abundance and length-age composition of the haddock stock in different parts of the bank. New survey indexes from whole area will be used for the assessment once the time-series for the whole area of haddock distribution is of sufficient length.
7. Analysis of possibility improving relationship between the survey assessed 0-group and the recruitment-at-age 1 assessed by VPA need for the short-term forecast.
8. There are doubts on the level of agreement of age reading by international experts.
9. Finally, it would be beneficial to develop and introduce standardization methods for reading the age for haddock.

### 7.9 Management considerations

The new FmSY estimate is consistent with the $F$ in the management plan previously evaluated by $^{\text {m }}$ ICES. The stock appears to be recovering after a period of very low recruitment. Incoming recruitment is still not as strong as it was historically. So a sudden expansion of the fishery at Rockall should be avoided.

A discards ban has been in place in the NEAFC regulatory area since 2009. Haddock in $6 . b$ have not yet been included under the EU landings obligation in 2016 (EC, 2015). However, the discard rate has not changed significantly and remains at a high level. Since 2017, basis for the ICES advice for haddock $6 . \mathrm{b}$ was changed from landings advice to catch advice as a result of a high level of discards not being accounted for. As result, ICES advice has become increasingly. It would be beneficial to develop and introduce measures aimed at preventing discards of haddock into fisheries practice. Elaboration of such measures complies with recommendations under the UNGA Resolution 61/105 that urges states to take action to reduce or eliminate fish discards (UNGA Resolution 61/105, 2007, Chapter VIII, item 60).

### 7.10 References

Blacker R.W. 1982. Rockall and its fishery. Laboratory Leaflet, Lowestoft. No 55.23 pp.
Brodziak J. 2005. Technical Description of STATCAM Version 1.2. Northeast Fisheries Science Center 166 Water Street Woods Hole, MA 02543. 39 pp.
Chuksin, Yu. V. and Gerber, E. M. 1976. Soviet fishery in the Rockall and Porcupine areas. Zaprybpromrazvedka. Kaliningrad, 8 pp.
EC. 2015. Commission Delegated Regulation (EU) 2015/2438 of 12 October 2015 establishing a discard plan for certain demersal fisheries in north-western waters

Finina E.A., Khlivnoy V.N., VinnichenkoV.I. 2009. The Reproductive Biology of Haddock (Mellanogrammus aeglefinus) at the Rockall Bank. Journal of Northwest Atlantic Fishery Science, Vol. 40: pp. 59-73.

ICES. 2004. Report of an Expert Group on Rockall Haddock Recovery Plans following a request for advice made on behalf of the European Community and the Russian Federation. 13-15 January 2004. Galway, Ireland. ICES/ACFM. 300 p .
ICES. 2013. ICES Advice 5.3.2.2. Special request Advice August 2013. Request from NEAFC to evaluate the proposals for the harvest control components of the management plan for Rockall haddock fisheries. ICES CM 2013/ACOM. 8 p.

ICES. 2016. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4). ICES CM 2015/ACOM:58. 183 p.

ICES. 2019. Report of the benchmark workshop (WKROCK) for Rockall haddock. ICES CM 2017/ACOM:58. 61 p.
Khlivnoy V.N., Sentiabov E.V. 2009. Influence of environmental factors on formation of year classes of haddock (Melanogrammusaeglefinus) at the Rockall Bank. Theses and reports of X All-Russian conference on problems of fishery forecasting, Murmansk: PINRO. pp. 137.

Newton, A. W., Peach, K. J., Coull, K. A., Gault, M. and Needle, C. L. 2008. Rockall and the Scottish haddock fishery. Fisheries Research, 94: pp.133-140.

Shestov, V. P. 1977. Rockall haddock. Fishery biological resources of the North Atlantic and adjacent seas of the Arctic Ocean. Moscow, pp.344-346.

Sonina M.A. 1976. The condition of the Arcto-Norwegian haddock stock and the factors determining the population size. // The Edition of the PINRO works. Vol. 37. pp. 129-150.
Tormosova I.D. 1978. The survival of North Sea haddock eggs at different stages of development and its determining factors. // The Edition of the AtlantNIRO works. Vol. 81. pp. 7-18.

## Table 4.3.1. Nominal catch (tonnes) of haddock in Division VIb, 1998-2018, as officially reported to ICES.

| Country | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | - | - | $\mathrm{n} / \mathrm{a}$ | n/a | - | - | - | - | 2 | 2 | 16 | - | 42 | 2 | 53 | - | <1 | <1 | - |  |  |
| France | - |  | 5 | 2 | - | 1 | - | - | - | - | - | - | - | <1 | - | - | <1 | - | - |  |  |
| Iceland | - | 167 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |
| Ireland | 704 | 1,021 | 824 | 357 | 206 | 169 | 19 | 105 | 41 | 338 | 721 | 352 | 169 | 123 | 31 | 105 | 94 | 190 | 362 | 500 | 433 |
| Norway | 40 | 61 | 152 | 70 | 49 | 60 | 32 | 33 | 123 | 84 | 36 | 71 | 65 | 40 | 48 | 121 | 41 | 66 | 63 | 26 | 16 |
| Portugal | 4 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |  |
| Russian Federation | - | 458 | 2,154 | 630 | 1,630 | 4,237 | 5,844 | 4,708 | 2,154 | 1,282 | 1669 | 55 | 198 | - | 1 | 4 | 388 | 136 | - | 153 | - |
| Spain | 21 | 25 | 47 | 51 | 7 | 19 | - | - | 5 | - | - | - | - | - | - | - | - | - | - | - |  |
| UK (E, W \& NI) | 561 | 288 | 36 | - | - | 56 | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |
| UK (Scotland) | 3,768 | 3,970 | 2,470 | 1,205 | 1,145 ${ }^{3}$ | 1,607 | $411^{3}$ | $332{ }^{3}$ | $440^{3}$ | 1,643 ${ }^{3}$ | 1,779 ${ }^{3}$ | 2,951 ${ }^{3}$ | 2,931 ${ }^{3}$ | 1,738 ${ }^{3}$ | $577{ }^{3}$ | 5963 | $1,152^{3}$ | 2,052 ${ }^{3}$ | $2,160^{3}$ | 3,930 ${ }^{3}$ | 3,418 |
| Total | 5,098 | 5,990 | 5,688 | 2,315 | 3,037 | 6,148 | 6,306 | 5,178 | 2,765 | 3,349 | 4,221 | 3,429 | 3,405 | 1,903 | 710 | 826 | 1,675 | 2,445 | 2,585 | 4,610 | 3,868 |
| Unallocated catch | -599 | -851 | -357 | -279 | 299 | 94 | 139 | 1 | 0 | 0 | 0 | -192 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| WG estimate | 4,499 | 5,139 | 5,331 ${ }^{4}$ | 2,0364 | 3,3364 | $6.242^{4}$ | 6,445 | 5,179 | 2,765 | 3,349 | 4,221 | 3,237 | 3,405 | 1,903 | 710 | 826 | 1,675 | 2,445 | 2,585 | 4,610 | 3,868 |

${ }^{1}$ Preliminary.
${ }^{2}$ Included in Division 6.a.
${ }^{3}$ Includes Scotland, England, Wales and NI landings
${ }^{4}$ Includes the total Russian catch.
$n / \mathbf{a}=$ not available.

Table 4.3.2. Details of Scottish and Irish effort (in hours) from 1985-2015 (preliminary data, no data for 20162018).

| Year | Scottish fleet |  |  | Irish fleet IROTB* |
| :---: | :---: | :---: | :---: | :---: |
|  | SCOTRL* | SCOLTR* | SCOSEI* |  |
| 1985 | 8421 | 3081 | 1677 |  |
| 1986 | 7465 | 4783 | 507 |  |
| 1987 | 8786 | 9737 | 402 |  |
| 1988 | 12450 | 5521 | 261 |  |
| 1989 | 10161 | 11946 | 1411 |  |
| 1990 | 3249 | 5335 | 4552 |  |
| 1991 | 2995 | 11464 | 6733 |  |
| 1992 | 2402 | 9623 | 3948 |  |
| 1993 | 1632 | 11540 | 1756 |  |
| 1994 | 2305 | 15543 | 399 |  |
| 1995 | 1789 | 13517 | 1383 | 9142 |
| 1996 | 1627 | 17324 | 952 | 7219 |
| 1997 | 563 | 16096 | 1061 | 7169 |
| 1998 | 1332 | 12263 | 456 | 7461 |
| 1999 | 11336 | 9424 | 456 | 8680 |
| 2000 | 12951 | 8586 | 80 | 9883 |
| 2001 | 7838 | 1037 | 42 | 7244 |
| 2002 | 8304 | 1100 | 0 | 2626 |
| 2003 | 15000 | 500 | 50 | 4618 |
| 2004 | 15200 | 300 | 50 | 2070 |
| 2005 | 7788 | 32 | 0 | 2693 |
| 2006 | 9990 | 231 | 0 | 5903 |
| 2007 | 4534 | 319 | 44 | 6589 |
| 2008 | 2497 | 1016 | 82 | 9740 |
| 2009 | NA | NA | NA | 4354 |
| 2010 | NA | NA | NA | 3280 |
| 2011 | NA | NA | NA | 2495 |
| 2012 | NA | NA | NA | 3291 |
| 2013 | NA | NA | NA | 2947 |
| 2014 | NA | NA | NA | 3159 |
| 2015 | NA | NA | NA | 3053 |
| 2016 | NA | NA | NA | NA |
| 2017 | NA | NA | NA | NA |
| 2018 | NA | NA | NA | NA |

SCOTRL* - Scottish Heavy Trawl, SCOLTR* - Scottish Light Trawl, SCOSEI* - Scottish Seine, IROTB* - Irish bottom otter trawl.

Table 4.3.3. Effort from the Scottish TR1 fleet and Irish otter-trawl fleet at Rockall (see the Section Cod VIb).

| year | Scottish TR1 fleet effort(kwdays) | Irish otter-trawl fleet Effort '000s Hrs |
| :---: | :---: | :---: |
| 2003 | 2504466 | 4.542 |
| 2004 | 1842103 | 2.233 |
| 2005 | 1217357 | 3.283 |
| 2006 | 1011354 | 5.9 |
| 2007 | 1060551 | 6.587 |
| 2008 | 1124197 | 9.898 |
| 2009 | 1631239 | 4.353 |
| 2010 | 1744452 | 3.28 |
| 2011 | 1565753 | 2.534 |
| 2012 | 901552 | 3.248 |
| 2013 | 532767 | 3.809 |
| 2014 | 668665 | 4.2 |
| 2015 | 563098 | 4.7 |
| 2016 | 514486 | 6.2 |
| 2017 | 794571 | 14.9 |
| 2018 | NA | NA |

Table 4.3.4. Discards and retained catches of haddock (number per trip) by Irish discard trips in the Rockall area from 2007-2009 and 2011-2012.

| Year | 2007 |  | 2008 |  | 2009 |  | 2011 |  | 2012 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Length (cm) | Discards | Retained <br> Catch | Discards | Retained <br> Catch | Discards | Retained <br> Catch | Discards | Re- <br> tained <br> Catch | Discards | Re- <br> tained <br> Catch |
| 10 |  |  |  |  |  |  |  |  | 1 |  |
| 11 |  |  |  |  |  |  |  |  | 1 |  |
| 12 |  |  |  |  |  |  |  |  | 1 |  |
| 13 |  |  |  |  |  |  |  |  | 1 |  |
| 14 |  |  |  |  |  |  |  |  |  |  |
| 15 |  |  |  |  |  |  |  |  |  |  |
| 16 |  |  |  |  |  |  |  |  |  |  |
| 17 |  |  |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |  |  |  |
| 19 | 1.3 |  |  |  |  |  |  |  |  |  |
| 20 |  |  |  |  |  |  |  |  |  |  |
| 21 |  |  |  |  |  |  |  |  |  |  |
| 22 | 1.6 |  | 14.8 |  |  |  |  |  |  |  |
| 23 | 4.6 |  | 66.2 |  |  |  | 13.1 |  |  |  |
| 24 | 7.3 |  | 183.8 |  |  |  | 98.9 | 5.7 |  |  |
| 25 | 22.7 |  | 576.9 |  | 15.6 |  | 53.9 | 5.7 |  |  |
| 26 | 54.2 |  | 1424.9 |  | 30.4 |  | 75.3 | 11.4 |  |  |
| 27 | 104.6 |  | 3024.6 |  | 25.2 |  | 121.3 | 34.3 | 2 |  |
| 28 | 256.9 |  | 6274.7 |  | 228.2 |  | 96.4 | 108.5 |  |  |
| 29 | 386.5 | 7.9 | 7193.3 |  | 180.6 |  | 33.6 | 62.8 |  |  |
| 30 | 533.4 | 17.6 | 7813.5 | 13.9 | 573.2 | 9.9 | 73.9 | 5.7 | 3 | 2 |
| 31 | 462.6 | 47.2 | 7573.7 | 40.6 | 1338.1 | 9.9 | 28.6 | 17.1 | 6 | 3 |
| 32 | 298.8 | 88.3 | 4639.0 | 77.8 | 1762.8 | 57.8 | 46.9 | 125.3 | 7 | 4 |
| 33 | 227.3 | 99.4 | 3664.7 | 126.8 | 2256.5 | 235.9 | 20.7 | 92.4 | 9 | 5 |
| 34 | 120.8 | 139.2 | 2391.8 | 277.4 | 1496.5 | 397.3 | 16.0 | 196.8 | 7 | 7 |
| 35 | 78.3 | 118.8 | 1590.1 | 503.6 | 656.6 | 614.8 | 4.8 | 118.6 | 6 | 8 |


| Year | 2007 |  | 2008 |  | 2009 |  | 2011 |  | 2012 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Length (cm) | Discards | Retained Catch | Discards | Retained Catch | Discards | Retained Catch | Discards | Retained Catch | Discards | Re- <br> tained Catch |
| 36 | 27.4 | 187.0 | 871.7 | 580.5 | 423.5 | 567.1 | 0.3 | 340.4 | 2 | 6 |
| 37 | 26.1 | 139.8 | 280.3 | 640.9 | 66.9 | 526.8 | 0.0 | 235.8 | 1 | 11 |
| 38 | 24.3 | 142.7 | 78.3 | 581.9 | 57.4 | 421.4 | 0.0 | 632.2 |  | 8 |
| 39 | 3.4 | 162.5 | 206.6 | 443.0 | 23.1 | 346.9 | 4.8 | 312.7 |  | 11 |
| 40 | 8.7 | 119.4 | 37.5 | 535.6 |  | 281.4 |  | 158.9 |  | 9 |
| 41 | 1.3 | 133.8 | 5.2 | 310.7 |  | 197.9 |  | 203.4 |  | 12 |
| 42 | 4.6 | 133.1 | 5.2 | 334.7 |  | 155.7 |  | 348.1 |  | 13 |
| 43 | 3.2 | 109.3 |  | 333.5 |  | 195.1 |  | 225.4 |  | 11 |
| 44 |  | 118.6 |  | 291.1 |  | 201.7 |  | 305.4 |  | 13 |
| 45 |  | 97.9 |  | 253.6 |  | 149.9 |  | 226.0 |  | 10 |
| >45 cm |  | 574.5 | 0.0 | 1791.2 | 0.0 | 1001.7 |  | 2490.8 | 1 | 144 |
| Total | 2659.9 | 2436.9 | 47916.8 | 7136.8 | 9134.4 | 5371.3 | 688.6 | 6263.7 | 48.0 | 277.0 |
| Discard rate, \% | 52.2 |  | 87.0 |  | 63.0 |  | 10.0 |  | 14.8 |  |

Table 4.3.5. Length composition of Irish discards and landings of haddock (number) by results of Irish discard trips in the Rockall area in 2014-2015.

| Year | 2014 |  | 2015 |  |
| :---: | :---: | :---: | :---: | :---: |
| Length (cm) | Discards | Landings | Discards | Landings |
| 10 |  |  |  |  |
| 11 |  |  |  |  |
| 12 |  |  |  |  |
| 13 |  |  |  |  |
| 14 |  |  |  |  |
| 15 |  |  |  |  |
| 16 |  |  |  |  |
| 17 |  |  |  |  |
| 18 |  |  |  |  |
| 19 |  |  |  |  |
| 20 | 508.86 |  |  |  |
| 21 | 1249.21 |  | 68.03 |  |
| 22 | 3757.56 |  | 136.45 |  |
| 23 | 9882.93 |  | 548.57 |  |
| 24 | 17742.15 |  | 2466.15 |  |
| 25 | 26690.88 |  | 5489.88 |  |
| 26 | 29456.22 | 206.22 | 8664.85 |  |
| 27 | 27737.04 | 1787.22 | 17011.27 |  |
| 28 | 28506.24 | 4605.52 | 23581.32 |  |
| 29 | 23556.01 | 5224.18 | 28730.09 |  |
| 30 | 22791.88 | 4261.83 | 33689.11 | 274.85 |
| 31 | 25734.19 | 4330.57 | 32838.74 | 742.11 |
| 32 | 25404.86 | 3436.96 | 33210.44 | 1044.45 |
| 33 | 17211.02 | 4880.48 | 25934.47 | 2308.78 |
| 34 | 8877.72 | 6392.74 | 17534.75 | 2666.09 |
| 35 | 4733.26 | 7217.61 | 7589.53 | 8300.60 |
| 36 | 2034.38 | 6324.00 | 4142.17 | 9702.36 |


| Year | 2014 |  | 2015 |  |
| :---: | :---: | :---: | :---: | :---: |
| Length (cm) | Discards | Landings | Discards | Landings |
| 37 | 918.99 | 5774.09 | 854.19 | 16628.69 |
| 38 | 77.02 | 4674.26 | 110.53 | 10636.86 |
| 39 | 153.20 | 3780.65 | 88.60 | 13495.35 |
| 40 | 0.00 | 4949.22 |  | 14787.16 |
| 41 | 39.00 | 4949.22 |  | 12808.21 |
| 42 | 51.67 | 7011.39 |  | 17425.77 |
| 43 | 12.67 | 4743.00 |  | 14732.19 |
| 44 | 12.67 | 4055.61 |  | 11488.91 |
| 45 | 25.34 | 2680.83 |  | 11186.57 |
| >45 cm | 290.53 | 30520.19 |  | 77254.68 |
| Total | 277455.52 | 121805.80 | 242689.10 | 225483.63 |
| Discard rate, \% | 69.5 |  | 51.8 |  |

Table 4.3.6. Discards and retained catches of haddock (number per trip) by Scottish discard trips in the Rockall area in 2009 and 2011-2015.

| Length (cm) | 2009 |  | 2011 |  | 2012 |  | 2013* |  | 2014* |  | 2015* |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings |
| 9 |  |  |  |  | 1.0 |  |  |  |  |  |  |  |
| 10 |  |  |  |  | 3.0 |  |  |  |  |  |  |  |
| 11 |  |  |  |  | 5.2 |  |  |  |  |  |  |  |
| 12 |  |  |  |  | 66.5 |  |  |  |  |  |  |  |
| 13 |  |  |  |  | 233.3 |  |  |  |  |  |  |  |
| 14 |  |  |  |  | 313.0 |  |  |  |  |  |  |  |
| 15 |  |  |  |  | 842.8 |  |  |  |  |  |  |  |
| 16 |  |  |  |  | 516.7 |  | 226 |  | 1493 |  |  |  |
| 17 |  |  |  |  | 247.3 |  | 0 |  | 7817 |  | 138 |  |
| 18 |  |  |  |  | 341.7 |  | 0 |  | 22709 |  | 957 |  |
| 19 |  |  |  |  | 81.5 |  | 135 |  | 39126 |  | 4591 |  |
| 20 |  |  |  |  | 4.7 |  | 39 |  | 37513 |  | 9278 |  |
| 21 |  |  |  |  |  |  | 357 |  | 25979 |  | 15194 |  |
| 22 |  |  |  |  |  |  | 1322 |  | 8774 |  | 16591 |  |
| 23 |  |  |  |  | 4.0 |  | 2201 |  | 14104 |  | 19529 |  |


| Length (cm) | 2009 |  | 2011 |  | 2012 |  | 2013* |  | 2014* |  | 2015* |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings |
| 24 |  |  |  |  | 23.0 |  | 3665 |  | 28818 |  | 42079 |  |
| 25 |  |  |  |  | 18.9 |  | 6643 |  | 64709 |  | 122065 |  |
| 26 |  |  | 3.8 |  | 36.4 |  | 6714 |  | 118616 |  | 206928 |  |
| 27 |  |  | 3.8 |  | 15.9 |  | 6424 |  | 164637 |  | 254254 |  |
| 28 | 24.2 |  | 17.4 |  | 22.6 |  | 5018 |  | 142534 |  | 305155 |  |
| 29 | 14.7 |  | 78.6 |  | 53.4 |  | 3599 |  | 121740 | 1422 | 342216 |  |
| 30 |  |  | 53.0 |  | 77.9 | 37.3 | 2326 |  | 78972 | 7965 | 330023 | 10543 |
| 31 | 5.3 | 26.4 | 17.4 |  | 126.6 | 76.1 | 1286 | 894 | 58592 | 25316 | 178402 | 31628 |
| 32 | 12.0 |  | 35.2 | 317.1 | 119.9 | 161.9 | 1181 | 2682 | 31670 | 30389 | 94018 | 84630 |
| 33 | 20.1 | 47.1 | 28.0 | 463.7 | 160.4 | 464.8 | 643 | 6454 | 13957 | 33340 | 23867 | 195299 |
| 34 |  | 201.7 |  | 637.4 | 71.0 | 1093.8 | 208 | 18902 | 10246 | 52890 | 9191 | 271402 |
| 35 |  | 220.2 | 139.8 | 1171.2 | 25.6 | 1366.4 | 101 | 23579 | 3404 | 47790 |  | 328955 |
| 36 |  | 269.0 | 139.8 | 1709.7 | 42.0 | 1872.7 | 39 | 34036 |  | 60976 |  | 241848 |
| 37 |  | 296.5 |  | 1668.7 | 10.1 | 2164.3 |  | 35748 |  | 57701 |  | 277221 |
| 38 |  | 353.1 | 139.8 | 2032.6 | 17.5 | 1917.5 |  | 33986 |  | 57472 |  | 197661 |
| 39 |  | 193.2 |  | 1927.7 |  | 2393.7 | 39 | 27892 |  | 61971 |  | 256136 |
| 40 |  | 237.9 | 139.8 | 1233.5 |  | 2091.6 |  | 36058 |  | 45808 |  | 188271 |


| Length (cm) | $2009$ <br> Discards | Landings | $2011$ <br> Discards | Landings | $2012$ <br> Discards | Landings | 2013* <br> Discards | Landings | 2014* <br> Discards | Landings | 2015* <br> Discards | Landings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| 41 |  | 131.7 |  | 1020.3 | 1.5 | 1876.3 |  | 23821 |  | 42575 |  | 189250 |
| 42 |  | 107.9 |  | 959.1 |  | 1247.9 |  | 18935 |  | 50824 |  | 123229 |
| 43 |  | 181.9 |  | 641.2 | 118.0 | 1416.8 |  | 23001 |  | 48330 |  | 150363 |
| 44 |  | 96.8 | 139.8 | 406.0 | 118.0 | 1288.2 |  | 20654 |  | 48019 |  | 108077 |
| 45 |  | 72.1 |  | 233.1 |  | 1326.8 |  | 22804 |  | 40359 |  | 75009 |
| 46 |  | 82.4 | 139.8 | 138.1 | 2.1 | 1252.9 |  | 22272 |  | 34162 |  | 78581 |
| 47 |  | 46.8 |  | 122.2 | 193.5 | 1023.0 |  | 22565 |  | 36909 |  | 39233 |
| 48 |  | 47.0 | 139.8 | 55.9 |  | 833.8 |  | 17565 |  | 33530 |  | 43136 |
| 49 |  | 33.3 | 1.0 | 49.9 | 194.5 | 711.7 |  | 18802 |  | 29220 |  | 48753 |
| 50 |  | 19.3 |  | 36.2 | 1.0 | 651.6 |  | 17499 |  | 28263 |  | 42833 |
| 51 |  | 8.9 |  | 37.5 |  | 410.3 |  | 12020 |  | 22682 |  | 50870 |
| 52 |  | 4.8 |  | 14.7 |  | 315.2 |  | 14866 |  | 23089 |  | 72142 |
| 53 |  | 5.1 |  | 20.5 |  | 206.1 |  | 12313 |  | 27292 |  | 40558 |
| 54 |  | 3.2 |  | 8.4 |  | 210.4 |  | 18722 |  | 34873 |  | 9895 |
| 55 |  | 2.3 |  | 5.4 |  | 98.8 | 26 | 11861 |  | 23816 |  | 34552 |
| 56 |  | 4.6 |  | 3.4 |  | 203.3 |  | 19573 |  | 18753 |  | 12660 |
| 57 |  | 2.7 |  | 1.6 |  | 408.4 |  | 14254 |  | 17896 |  | 9895 |


| Length (cm) | 2009 |  | 2011 |  | 2012 |  | 2013* |  | 2014* |  | 2015* |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings |
| 58 |  | 1.9 |  | 3.1 |  | 404.8 |  | 8962 |  | 16511 |  | 9506 |
| 59 |  | 1.7 |  | 9.1 |  | 87.8 |  | 6702 |  | 21930 |  | 7518 |
| 60 |  | 1.2 |  |  |  | 189.9 |  | 9813 |  | 20822 |  | 2765 |
| 61 |  | 1.7 |  | 2.7 |  | 190.7 |  | 5851 |  | 12248 |  |  |
| 62 |  | 1.1 |  | 1.3 |  | 213.7 |  | 6436 |  | 20519 |  | 5531 |
| 63 |  | 0.5 |  | 2.4 |  | 210.2 |  | 4016 |  | 9150 |  |  |
| 64 |  | 1.3 |  |  |  | 97.7 |  | 6675 |  | 7792 |  | 1166 |
| 65 |  |  |  | 1.1 |  | 45.1 |  | 5212 |  | 9321 |  |  |
| 66 |  |  |  | 1.1 |  | 105.2 |  | 2314 |  | 13225 |  |  |
| 67 |  |  |  |  |  | 45.0 |  | 3830 |  | 14393 |  |  |
| 68 |  |  |  | 1.0 |  | 24.3 |  | 1649 |  | 9712 |  | 3154 |
| 69 |  |  |  |  |  | 63.1 |  | 1649 |  | 3359 |  |  |
| 70 |  |  |  | 0.9 |  | 58.0 |  | 1915 |  | 4556 |  |  |
| 71 |  |  |  |  |  | 47.9 |  | 665 |  | 2406 |  |  |
| 72 |  |  |  |  |  | 42.2 |  | 1782 |  | 190 |  |  |
| 73 |  |  |  |  |  | 20.1 |  | 1117 |  | 1102 |  | 2765 |
| 74 |  |  |  |  |  | 20.6 |  | 133 |  | 2181 |  |  |


| Length (cm) | 2009 |  | 2011 |  | 2012 |  | 2013* |  | 2014* |  | 2015* |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings |
| 76 |  |  |  |  |  | 5.7 |  |  |  |  |  |  |
| 77 |  |  |  |  |  | 8.6 |  |  |  | 71 |  |  |
| 78 |  |  |  | 0.7 |  | 4.1 |  |  |  | 759 |  |  |
| 82 |  |  |  | 0.6 |  |  |  |  |  |  |  |  |
| Total | 76.3 | 2705.3 | 1216.8 | 14939.0 | 4110.5 | 29006.3 | 42218 | 600479 | 995410 | 1214092 | 1974476 | 3245035 |
| Discard rate, \% | 2.7 |  | 7.5 |  | 12.4 |  | 6.6 |  | 45.0 |  | 37.8 |  |

*Retained discards and landings

Table 4.3.7. Discards and retained catches of haddock (number) by Scottish and Irish discard trips in the Rockall area in 2013-2018.

| Year | Country |  | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 1 | 2 | 3 | 4 | 5 | 6 | $7+$ |
| 2013 | Scotland | Landings | 116013 | 9886 | 1154 | 33064 | 4373 | 33020 | 3387 |
|  |  | Discards | 4666330 | 28973 | 0 | 0 | 0 | 0 | 11791 |
|  | Ireland* | Landings | - | - | - | - | - | - | - |
|  |  | Discards | 55362 | 5189 | 9389 | 3816 | 31041 | 35875 | 0 |
|  | Ireland** | Landings | - | - | - | - | - | - | - |
|  |  | Discards | 3061 | 2869 | 5192 | 2110 | 1716 | 1984 | 0 |
| 2014 | Scotland | Landings | - | 577684 | 2252 | 213 | 87220 | 18169 | 528556 |
|  |  | Discards | 142263 | 853148 | - | - | - | - | - |
|  | Ireland | Landings | 4188 | 58642 | 2353 | 1277 | 21085 | 7630 | 26631 |
|  |  | Discards | 15651 | 261804 | - | - | - | - | - |
| 2015 | Scotland | Landings | - | 464407 | 2679182 | 1620 | 1171 | 24139 | 88332 |
|  |  | Discards | 70129 | 1935829 | 45431 | - | - | - | - |
|  | Ireland | Landings | - | 2277 | 159849 | 3767 | 3662 | 42685 | 13244 |
|  |  | Discards | - | 149261 | 93428 | - | - | - | - |
| 2016 | Scotland | Landings | 127 | 580 | 1991 | 590 | 0 | 0 | 2891 |
|  |  | BMS landings | 1271 | 356 | 51 | - | - | - |  |


| Year | Country |  | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 1 | 2 | 3 | 4 | 5 | 6 | $7+$ |
|  |  | Discards | 163346 | 153742 | 88894 | 402 | - | - | - |
|  | Ireland | Landings | - | 27955 | 138593 | 278405 | 3345 | 2294 | 8634 |
|  |  | BMS landings | - | - | - | - | - | - | - |
|  |  | Discards | 23629 | 177594 | 287589 | 108446 | - | - | - |
| 2017 | Scotland | Landings | 340 | 955346 | 1401088 | 1606845 | 821574 | 2851 | 12316 |
|  |  | BMS landings | - | - | - | - | - | - |  |
|  |  | Discards | 747839 | 245953 | 1073 | 201 | 268 | - | - |
|  | Ireland | Landings | 24 | 166140 | 75380 | 217982 | 125193 | 4364 | 9657 |
|  |  | BMS landings | - | - | - | - | - | - | - |
|  |  | Discards | 314743 | 43494 | 19349 | 12118 | - | - | - |
| 2018 | Scotland | Landings |  | 3116059 | 456039 | 2052985 | 533709 | 191175 | 8853 |
|  |  | BMS landings | - | - | - | - | - | - | - |
|  |  | Discards | 87472 | 2906183 | 2033 | 38342 | 458 | 431 |  |
|  | Ireland | Landings |  |  |  |  |  |  |  |
|  |  | BMS landings |  | 33562.58 | 6180.39 | 4416.73 | 17015.94 | 13023.1 | 3205.62 |
|  |  | Discards | 219.48 | 42390.14 | 790.51 | 1315.76 |  |  |  |

[^4]** Mesh size 70-99 mm.

Table 4.3.8. Haddock in 6.b. Tuning data available from the Scottish groundfish survey conducted in September. In bold, the data used in the assessment.

HADDOCK WGCSE 2017 ROCKALL
101
SCOGFS
19912018
110.660 .75

06

| 1 | 14458 | 16398 | 4431 | 683 | 315 | 228 | 37 | 64 | 3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 20336 | 44912 | 14631 | 3150 | 647 | 127 | 200 | 4 | 32 |
| 1 | 15220 | 37959 | 15689 | 3716 | 1104 | 183 | 38 | 73 | 21 |
| 1 | 23474 | 13287 | 11399 | 4314 | 969 | 203 | 30 | 12 | 4 |
| 1 | 16923 | 16971 | 6648 | 5993 | 1935 | 483 | 200 | 16 | -1 |
| 1 | 33578 | 19420 | 5903 | 1940 | 1317 | 325 | 69 | 6 | 1 |
| 1 | 28897 | 10693 | 2384 | 538 | 292 | 281 | 71 | 9 | 1 |
| 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1 | 10178 | 9969 | 2410 | 708 | 279 | 172 | 90 | 64 | 32 |
| 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1 | 31813 | 7455 | 521 | 284 | 154 | 39 | 14 | 12 | 14 |
| 1 | 11704 | 20925 | 2464 | 173 | 105 | 65 | 20 | 10 | 15 |
| 1 | 2526 | 10114 | 10927 | 1656 | 138 | 97 | 100 | 26 | 6 |
| 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1 | 24452 | 4082 | 920 | 1506 | 2107 | 231 | 33 | 13 | 7 |
| 1 | 3570 | 18715 | 2562 | 256 | 1402 | 1694 | 349 | 16 | 6 |
| 1 | 558 | 2671 | 6019 | 570 | 254 | 516 | 367 | 28 | 2 |
| 1 | 85 | 560 | 966 | 3813 | 182 | 41 | 282 | 249 | 49 |
| 1 | 132 | 139 | 323 | 488 | 1651 | 40 | 9 | 54 | 17 |
| 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1 | 13 | 17 | 96 | 22 | 42 | 88 | 607 | 4 | 4 |
| 1 | 39619 | 4 | 12 | 73 | 14 | 75 | 50 | 635 | 9 |
| 1 | 6035 | 14179 | 5 | 8 | 8 | 9 | 11 | 23 | 166 |
| 1 | 3044 | 7232 | 4692 | 5 | 0 | 13 | 0 | 11 | 10 |
| 1 | 1997 | 2908 | 5635 | 3357 | 0 | 0 | 16 | 2 | 20 |
| 1 | 67096 | 1576 | 1483 | 2064 | 1526 | 11 | 1 | 5 | 2 |
| 1 | 30130 | 29449 | 956 | 909 | 1389 | 663 | 5 | 1 | 2 |
| 1 | 10008 | 4170 | 10535 | 308 | 773 | 967 | 358 | 0 | 0 |

Table 4.3.9. Haddock in 6.b. International landings, discards and total catch.

| Year | Num (*1000) <br> Landings | Discards | Total Catch ${ }^{1}$ | Weight, tonnes <br> Landings | Discards | Total Catch ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |
| 1991 | 12302 | 65832 | 78134 | 5655 | 13229 | 18884 |
| 1992 | 11418 | 55964 | 67383 | 5320 | 11873 | 17192 |
| 1993 | 8767 | 44656 | 53423 | 4784 | 9856 | 14639 |
| 1994 | 11400 | 46628 | 58028 | 5733 | 11027 | 16761 |
| 1995 | 11784 | 35467 | 47251 | 5587 | 9170 | 14758 |
| 1996 | 14066 | 41506 | 55572 | 7075 | 9356 | 16432 |
| 1997 | 9965 | 26980 | 36945 | 5166 | 5893 | 11059 |
| 1998 | 9034 | 47831 | 56865 | 4984 | 10863 | 15847 |
| 1999 | 12931 | 52881 | 65812 | 5358 | 11065 | 16423 |
| 2000 | 16000 | 26033 | 42033 | 5445 | 6611 | 12056 |
| 2001 | 5069 | 9222 | 14291 | 2020 | 1536 | 3556 |
| 2002 | 11168 | 21899 | 33067 | 3116 | 4153 | 7269 |
| 2003 | 24542 | 25087 | 49629 | 5967 | 5521 | 11488 |
| 2004 | 22706 | 3989 | 26695 | 6437 | 883 | 7321 |
| 2005 | 19505 | 1877 | 21382 | 5238 | 505 | 5742 |
| 2006 | 9605 | 1667 | 11273 | 2756 | 386 | 3142 |
| 2007 | 8936 | 12300 | 21236 | 3348 | 2242 | 5590 |
| 2008 | 10209 | 7603 | 17812 | 4221 | 2104 | 6325 |
| 2009 | 6709 | 4765 | 11474 | 3242 | 1556 | 4798 |
| 2010 | 5264 | 3242 | 8506 | 3404 | 907 | 4311 |
| 2011 | 3082 | 248 | 3331 | 1861 | 152 | 2013 |
| 2012 | 631 | 49 | 680 | 686 | 26 | 712 |
| 2013 | 829 | 5039 | 5868 | 889 | 1065 | 1954 |
| 2014 | 3114 | 1634 | 4748 | 1845 | 332 | 2177 |
| 2015 | 4327 | 2397 | 6724 | 2510 | 554 | 3064 |
| 2016 | 3733 | 1333 | 5068 | 2504 | 401 | 2905 |
| 2017 | 6629 | 1552 | 8181 | 4431 | 379 | 4810 |
| 2018 | 6985 | 3087 | 10072 | 3850 | 788 | 4638 |

Table 4. 3.10. Haddock in 6.b. International catch (landings and discards) numbers (*103) at-age.

|  |  |  | Age |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 21186 | 33847 | 15189 | 5341 | 1704 | 346 | 522 |
| 1992 | 16084 | 24711 | 18584 | 5361 | 1761 | 676 | 206 |
| 1993 | 11178 | 19375 | 15494 | 4938 | 1617 | 461 | 359 |
| 1994 | 8170 | 20623 | 17868 | 8209 | 2449 | 476 | 232 |
| 1995 | 2749 | 9831 | 21584 | 9756 | 2464 | 787 | 79 |
| 1996 | 12096 | 18811 | 10911 | 9612 | 3299 | 751 | 92 |
| 1997 | 9957 | 10535 | 5388 | 4098 | 5002 | 1758 | 206 |
| 1998 | 14224 | 19807 | 10173 | 4763 | 3740 | 2767 | 1391 |
| 1999 | 17282 | 21949 | 12203 | 5499 | 3419 | 2684 | 2776 |
| 2000 | 8222 | 12581 | 10698 | 4917 | 2050 | 1498 | 2066 |
| 2001 | 7669 | 2013 | 1699 | 821 | 1041 | 477 | 570 |
| 2002 | 13363 | 11119 | 4537 | 2445 | 898 | 260 | 444 |
| 2003 | 6576 | 23606 | 14568 | 2065 | 1286 | 927 | 602 |
| 2004 | 932 | 4112 | 10282 | 9212 | 1386 | 296 | 474 |
| 2005 | 1061 | 3723 | 7420 | 8124 | 753 | 109 | 193 |
| 2006 | 2880 | 1475 | 1626 | 2414 | 2291 | 436 | 151 |
| 2007 | 1489 | 9829 | 3630 | 1514 | 2227 | 1827 | 720 |
| 2008 | 476 | 2207 | 11437 | 1291 | 507 | 964 | 930 |
| 2009 | 223 | 707 | 1237 | 8046 | 495 | 263 | 504 |
| 2010 | 152 | 534 | 1064 | 2087 | 4096 | 276 | 296 |
| 2011 | 4 | 59 | 75 | 183 | 181 | 2579 | 249 |
| 2012 | 5 | 6 | 144 | 58 | 3 | 35 | 428 |
| 2013 | 4896 | 98 | 101 | 86 | 39 | 84 | 565 |
| 2014 | 406 | 3008 | 418 | 52 | 138 | 47 | 679 |
| 2015 | 80 | 2973 | 3387 | 104 | 7 | 61 | 112 |
| 2016 | 374 | 1051 | 2639 | 988 | 3 | 2 | 11 |
| 2017 | 1194 | 1670 | 1802 | 2191 | 1207 | 58 | 59 |
| 2018 | 88 | 6373 | 504 | 2273 | 598 | 222 | 13 |

Table 4.3.11. Haddock in 6.b. International landings numbers $\left({ }^{*} 10^{3}\right)$ at-age.

|  |  |  | Age |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 87 | 6807 | 3011 | 1344 | 558 | 32 | 464 |
| 1992 | 86 | 3642 | 5623 | 964 | 580 | 364 | 160 |
| 1993 | 28 | 1919 | 4740 | 1157 | 489 | 144 | 290 |
| 1994 | 30 | 1160 | 5299 | 3665 | 1039 | 66 | 141 |
| 1995 | 1 | 146 | 5205 | 4791 | 1319 | 279 | 43 |
| 1996 | 2 | 5149 | 1861 | 4149 | 2347 | 473 | 85 |
| 1997 | 0 | 319 | 2102 | 2155 | 3658 | 1540 | 192 |
| 1998 | 4 | 392 | 1815 | 1340 | 1898 | 2284 | 1301 |
| 1999 | 245 | 2600 | 2994 | 1972 | 1228 | 1600 | 2291 |
| 2000 | 33 | 3446 | 5081 | 3006 | 1296 | 1176 | 1963 |
| 2001 | 402 | 994 | 1116 | 555 | 991 | 462 | 549 |
| 2002 | 657 | 2983 | 3998 | 2111 | 809 | 217 | 392 |
| 2003 | 920 | 8103 | 11010 | 1848 | 1189 | 879 | 593 |
| 2004 | 197 | 1765 | 9502 | 9119 | 1364 | 286 | 472 |
| 2005 | 887 | 2835 | 6866 | 7913 | 725 | 98 | 182 |
| 2006 | 2344 | 768 | 1290 | 2356 | 2269 | 428 | 150 |
| 2007 | 31 | 1220 | 2709 | 1074 | 1550 | 1634 | 719 |
| 2008 | 17 | 749 | 6191 | 1164 | 479 | 761 | 848 |
| 2009 | 5 | 11 | 244 | 5243 | 460 | 261 | 486 |
| 2010 | 0 | 71 | 196 | 352 | 4078 | 274 | 294 |
| 2011 | 2 | 23 | 71 | 177 | 181 | 2405 | 222 |
| 2012 | 0 | 0 | 134 | 51 | 0 | 35 | 410 |
| 2013 | 162 | 14 | 2 | 46 | 6 | 46 | 553 |
| 2014 | 226 | 1553 | 418 | 52 | 138 | 47 | 679 |
| 2015 | 9 | 820 | 3214 | 104 | 7 | 61 | 112 |
| 2016 | 127 | 612 | 2137 | 842 | 3 | 2 | 11 |
| 2017 | 7 | 1336 | 1783 | 2179 | 1207 | 58 | 59 |
| 2018 | 0 | 3418 | 502 | 2233 | 598 | 222 | 13 |

Table 4.3.12. Haddock in 6.b. International discards numbers ( ${ }^{*} 10^{3}$ ) at age.

| YEAR | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 21099 | 27040 | 12178 | 3998 | 1146 | 313 | 58 |
| 1992 | 15998 | 21069 | 12961 | 4397 | 1181 | 312 | 46 |
| 1993 | 11151 | 17456 | 10755 | 3781 | 1128 | 317 | 69 |
| 1994 | 8140 | 19464 | 12570 | 4545 | 1409 | 410 | 91 |
| 1995* | 2748 | 9685 | 16379 | 4965 | 1145 | 508 | 36 |
| 1996 | 12094 | 13662 | 9051 | 5463 | 952 | 278 | 7 |
| 1997* | 9957 | 10216 | 3286 | 1944 | 1344 | 218 | 15 |
| 1998* | 14220 | 19415 | 8357 | 3423 | 1842 | 483 | 91 |
| 1999* | 17037 | 19348 | 9209 | 3526 | 2191 | 1084 | 485 |
| 2000* | 8189 | 9136 | 5616 | 1912 | 755 | 322 | 103 |
| 2001* | 7268 | 1019 | 583 | 266 | 50 | 15 | 21 |
| 2002 | 12706 | 8136 | 539 | 334 | 89 | 43 | 51 |
| 2003 | 5655 | 15503 | 3558 | 217 | 97 | 48 | 8 |
| 2004 | 735 | 2346 | 781 | 93 | 22 | 10 | 2 |
| 2005 | 174 | 888 | 554 | 210 | 28 | 11 | 11 |
| 2006 | 536 | 707 | 336 | 58 | 22 | 8 | 1 |
| 2007 | 1458 | 8609 | 921 | 440 | 678 | 193 | 0 |
| 2008 | 458 | 1458 | 5246 | 128 | 28 | 203 | 82 |
| 2009 | 218 | 696 | 993 | 2803 | 35 | 2 | 18 |
| 2010* | 152 | 463 | 868 | 1736 | 19 | 2 | 2 |
| 2011* | 2 | 36 | 4 | 6 | 0 | 174 | 27 |
| 2012* | 5 | 6 | 10 | 7 | 3 | 0 | 18 |
| 2013* | 4733 | 84 | 99 | 40 | 33 | 38 | 12 |
| 2014* | 179 | 1454 | 0 | 0 | 0 | 0 | 0 |
| 2015* | 71 | 2153 | 173 | 0 | 0 | 0 | 0 |
| 2016* | 245 | 439 | 503 | 146 | 0 | 0 | 0 |
| 2017* | 1187 | 334 | 20 | 12 | 0 | 0 | 0 |
| 2018* | 88 | 2955 | 3 | 40 | 0 | 0 | 0 |

[^5]Table 4.3.13. Haddock in 6.b. International catch (landings and discards) weights-at-age (kg).

| YEAR | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 0.142 | 0.240 | 0.291 | 0.378 | 0.469 | 0.414 | 0.681 |
| 1992 | 0.133 | 0.239 | 0.318 | 0.362 | 0.423 | 0.567 | 0.852 |
| 1993 | 0.137 | 0.238 | 0.335 | 0.400 | 0.493 | 0.503 | 0.882 |
| 1994 | 0.153 | 0.233 | 0.319 | 0.420 | 0.469 | 0.477 | 0.740 |
| 1995 | 0.118 | 0.222 | 0.309 | 0.401 | 0.501 | 0.460 | 0.870 |
| 1996 | 0.136 | 0.278 | 0.314 | 0.396 | 0.553 | 0.575 | 0.762 |
| 1997 | 0.136 | 0.240 | 0.322 | 0.381 | 0.512 | 0.634 | 0.940 |
| 1998 | 0.141 | 0.250 | 0.308 | 0.354 | 0.436 | 0.546 | 0.663 |
| 1999 | 0.138 | 0.208 | 0.272 | 0.334 | 0.379 | 0.483 | 0.619 |
| 2000 | 0.189 | 0.250 | 0.267 | 0.321 | 0.382 | 0.451 | 0.709 |
| 2001 | 0.133 | 0.264 | 0.326 | 0.447 | 0.427 | 0.520 | 0.683 |
| 2002 | 0.135 | 0.239 | 0.237 | 0.325 | 0.509 | 0.579 | 0.755 |
| 2003 | 0.153 | 0.203 | 0.256 | 0.349 | 0.384 | 0.424 | 0.604 |
| 2004 | 0.147 | 0.198 | 0.244 | 0.294 | 0.444 | 0.609 | 0.753 |
| 2005 | 0.114 | 0.197 | 0.235 | 0.311 | 0.459 | 0.600 | 1.062 |
| 2006 | 0.093 | 0.198 | 0.245 | 0.329 | 0.441 | 0.595 | 0.787 |
| 2007 | 0.114 | 0.186 | 0.265 | 0.294 | 0.386 | 0.496 | 0.578 |
| 2008 | 0.199 | 0.241 | 0.291 | 0.437 | 0.571 | 0.669 | 0.937 |
| 2009 | 0.248 | 0.288 | 0.339 | 0.391 | 0.668 | 0.513 | 1.012 |
| 2010 | 0.141 | 0.247 | 0.333 | 0.327 | 0.590 | 0.977 | 1.464 |
| 2011 | 0.198 | 0.280 | 0.596 | 0.449 | 0.695 | 0.603 | 0.748 |
| 2012 | 0.263 | 0.295 | 0.622 | 0.784 | 0.372 | 1.411 | 1.219 |
| 2013 | 0.211 | 0.368 | 0.236 | 0.704 | 0.423 | 0.827 | 1.261 |
| 2014 | 0.140 | 0.286 | 0.268 | 0.545 | 1.000 | 1.036 | 1.370 |
| 2015 | 0.104 | 0.254 | 0.601 | 0.354 | 1.178 | 0.948 | 1.439 |
| 2016 | 0.298 | 0.449 | 0.600 | 0.711 | 1.556 | 1.808 | 2.650 |
| 2017 | 0.219 | 0.430 | 0.586 | 0.691 | 0.944 | 0.780 | 1.270 |
| 2018 | 0.088 | 0.298 | 0.563 | 0.700 | 0.935 | 1.233 | 1.928 |

Table 4.3.14. Haddock in 6.b. International landings weights-at-age (kg).

| YEAR | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 0.302 | 0.402 | 0.444 | 0.592 | 0.724 | 0.963 | 0.704 |
| 1992 | 0.136 | 0.366 | 0.455 | 0.658 | 0.612 | 0.759 | 0.954 |
| 1993 | 0.305 | 0.402 | 0.503 | 0.701 | 0.830 | 0.820 | 0.972 |
| 1994 | 0.314 | 0.356 | 0.452 | 0.558 | 0.638 | 1.224 | 0.890 |
| 1995 | 0.377 | 0.311 | 0.414 | 0.479 | 0.640 | 0.699 | 1.236 |
| 1996 | 0.327 | 0.436 | 0.501 | 0.487 | 0.627 | 0.709 | 0.783 |
| 1997 | 0.300 | 0.315 | 0.401 | 0.444 | 0.564 | 0.661 | 0.973 |
| 1998 | 0.256 | 0.344 | 0.494 | 0.517 | 0.542 | 0.591 | 0.678 |
| 1999 | 0.085 | 0.177 | 0.326 | 0.417 | 0.495 | 0.595 | 0.662 |
| 2000 | 0.111 | 0.206 | 0.242 | 0.328 | 0.413 | 0.483 | 0.720 |
| 2001 | 0.094 | 0.281 | 0.344 | 0.497 | 0.427 | 0.522 | 0.690 |
| 2002 | 0.107 | 0.196 | 0.227 | 0.323 | 0.521 | 0.627 | 0.804 |
| 2003 | 0.100 | 0.164 | 0.246 | 0.350 | 0.387 | 0.423 | 0.606 |
| 2004 | 0.142 | 0.172 | 0.241 | 0.293 | 0.446 | 0.617 | 0.754 |
| 2005 | 0.103 | 0.184 | 0.230 | 0.310 | 0.461 | 0.614 | 1.095 |
| 2006 | 0.084 | 0.167 | 0.223 | 0.327 | 0.440 | 0.598 | 0.789 |
| 2007 | 0.096 | 0.238 | 0.275 | 0.322 | 0.449 | 0.521 | 0.578 |
| 2008 | 0.125 | 0.197 | 0.302 | 0.444 | 0.583 | 0.752 | 0.984 |
| 2009 | 0.300 | 0.346 | 0.420 | 0.416 | 0.692 | 0.512 | 1.031 |
| 2010 | 0.052 | 0.420 | 0.517 | 0.457 | 0.591 | 0.980 | 1.473 |
| 2011 | 0.214 | 0.329 | 0.613 | 0.454 | 0.694 | 0.594 | 0.780 |
| 2012 | 0.189 | 0.368 | 0.632 | 0.850 | 0.898 | 1.412 | 1.238 |
| 2013 | 0.510 | 0.554 | 0.713 | 0.972 | 1.361 | 0.948 | 1.267 |
| 2014 | 0.186 | 0.351 | 0.268 | 0.545 | 1.000 | 1.036 | 1.370 |
| 2015 | 0.107 | 0.327 | 0.615 | 0.354 | 1.178 | 0.948 | 1.439 |
| 2016 | 0.409 | 0.574 | 0.664 | 0.767 | 1.576 | 1.808 | 2.650 |
| 2017 | 0.173 | 0.460 | 0.587 | 0.692 | 0.944 | 0.780 | 1.270 |
| 2018 | -1 | 0.332 | 0.564 | 0.705 | 0.935 | 1.235 | 1.928 |

Table 4.3.15. Haddock in 6.b. International discards weights-at-age (kg).

| YEAR | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 0.142 | 0.199 | 0.253 | 0.306 | 0.345 | 0.358 | 0.499 |
| 1992 | 0.133 | 0.217 | 0.258 | 0.298 | 0.330 | 0.342 | 0.499 |
| 1993 | 0.137 | 0.220 | 0.260 | 0.307 | 0.346 | 0.359 | 0.504 |
| 1994 | 0.153 | 0.226 | 0.263 | 0.308 | 0.345 | 0.356 | 0.508 |
| 1995 | 0.118 | 0.220 | 0.276 | 0.325 | 0.341 | 0.329 | 0.438 |
| 1996 | 0.136 | 0.218 | 0.276 | 0.326 | 0.370 | 0.348 | 0.515 |
| 1997 | 0.136 | 0.238 | 0.272 | 0.312 | 0.372 | 0.442 | 0.512 |
| 1998 | 0.141 | 0.248 | 0.267 | 0.291 | 0.327 | 0.336 | 0.451 |
| 1999 | 0.139 | 0.212 | 0.255 | 0.288 | 0.313 | 0.318 | 0.417 |
| 2000 | 0.189 | 0.267 | 0.289 | 0.311 | 0.330 | 0.334 | 0.484 |
| 2001 | 0.135 | 0.247 | 0.294 | 0.344 | 0.412 | 0.440 | 0.513 |
| 2002 | 0.137 | 0.254 | 0.308 | 0.335 | 0.398 | 0.338 | 0.382 |
| 2003 | 0.161 | 0.223 | 0.287 | 0.342 | 0.337 | 0.440 | 0.487 |
| 2004 | 0.148 | 0.218 | 0.282 | 0.343 | 0.324 | 0.371 | 0.449 |
| 2005 | 0.171 | 0.240 | 0.298 | 0.357 | 0.387 | 0.473 | 0.511 |
| 2006 | 0.132 | 0.233 | 0.334 | 0.420 | 0.495 | 0.435 | 0.423 |
| 2007 | 0.115 | 0.179 | 0.233 | 0.227 | 0.243 | 0.280 | 0.420 |
| 2008 | 0.202 | 0.264 | 0.279 | 0.370 | 0.351 | 0.358 | 0.446 |
| 2009 | 0.247 | 0.287 | 0.319 | 0.343 | 0.360 | 0.662 | 0.507 |
| 2010 | 0.141 | 0.220 | 0.292 | 0.301 | 0.322 | 0.534 | 0.250 |
| 2011 | 0.178 | 0.248 | 0.300 | 0.302 | 0.795 | 0.727 | 0.481 |
| 2012 | 0.263 | 0.295 | 0.488 | 0.319 | 0.339 | 0.733 | 0.797 |
| 2013 | 0.201 | 0.337 | 0.228 | 0.397 | 0.247 | 0.679 | 0.980 |
| 2014 | 0.082 | 0.218 | - | - | - | - | - |
| 2015 | 0.104 | 0.227 | 0.334 | - | - | - | - |
| 2016 | 0.240 | 0.276 | 0.325 | 0.393 | - | - | - |
| 2017 |  | 0.308 | 0.482 | 0.520 | 0.726 | - | - |
| 2018 | 0.088 | 0.258 | 0.361 | 0.422 | 0.479 | 0.536 | -1 |

Table 4.3.16. Haddock 6.b. Stock weights-at-age (kg).

| YEAR | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 0.142 | 0.240 | 0.291 | 0.378 | 0.469 | 0.414 | 0.681 |
| 1992 | 0.133 | 0.239 | 0.318 | 0.362 | 0.423 | 0.567 | 0.852 |
| 1993 | 0.137 | 0.238 | 0.335 | 0.400 | 0.493 | 0.503 | 0.882 |
| 1994 | 0.153 | 0.233 | 0.319 | 0.420 | 0.469 | 0.477 | 0.740 |
| 1995 | 0.137 | 0.234 | 0.314 | 0.392 | 0.471 | 0.484 | 0.805 |
| 1996 | 0.136 | 0.242 | 0.319 | 0.396 | 0.488 | 0.516 | 0.821 |
| 1997 | 0.136 | 0.242 | 0.320 | 0.399 | 0.506 | 0.530 | 0.839 |
| 1998 | 0.137 | 0.245 | 0.314 | 0.390 | 0.494 | 0.538 | 0.795 |
| 1999 | 0.134 | 0.240 | 0.305 | 0.373 | 0.476 | 0.540 | 0.771 |
| 2000 | 0.148 | 0.245 | 0.297 | 0.357 | 0.452 | 0.538 | 0.739 |
| 2001 | 0.148 | 0.242 | 0.299 | 0.368 | 0.427 | 0.527 | 0.723 |
| 2002 | 0.147 | 0.242 | 0.282 | 0.356 | 0.426 | 0.516 | 0.686 |
| 2003 | 0.150 | 0.233 | 0.272 | 0.355 | 0.416 | 0.491 | 0.674 |
| 2004 | 0.151 | 0.231 | 0.266 | 0.347 | 0.429 | 0.517 | 0.701 |
| 2005 | 0.136 | 0.220 | 0.260 | 0.345 | 0.444 | 0.546 | 0.771 |
| 2006 | 0.128 | 0.207 | 0.243 | 0.322 | 0.447 | 0.561 | 0.792 |
| 2007 | 0.124 | 0.197 | 0.249 | 0.315 | 0.423 | 0.545 | 0.757 |
| 2008 | 0.134 | 0.204 | 0.256 | 0.333 | 0.460 | 0.594 | 0.823 |
| 2009 | 0.154 | 0.222 | 0.275 | 0.352 | 0.505 | 0.574 | 0.875 |
| 2010 | 0.159 | 0.232 | 0.295 | 0.355 | 0.531 | 0.650 | 0.956 |
| 2011 | 0.180 | 0.248 | 0.365 | 0.380 | 0.582 | 0.651 | 0.948 |
| 2012 | 0.210 | 0.270 | 0.436 | 0.477 | 0.579 | 0.834 | 1.076 |
| 2013 | 0.212 | 0.295 | 0.425 | 0.531 | 0.550 | 0.866 | 1.141 |
| 2014 | 0.191 | 0.295 | 0.411 | 0.562 | 0.616 | 0.971 | 1.212 |
| 2015 | 0.183 | 0.297 | 0.465 | 0.567 | 0.734 | 0.965 | 1.207 |
| 2016 | 0.203 | 0.330 | 0.465 | 0.619 | 0.906 | 1.206 | 1.588 |
| 2017 | 0.195 | 0.357 | 0.458 | 0.601 | 1.020 | 1.080 | 1.598 |
| 2018 | 0.170 | 0.343 | 0.524 | 0.600 | 1.123 | 1.161 | 1.731 |

Table 4.3.17. XSA diagnostics from the assessment of Haddock in 6.b. Final runs.


Table 4.3.17. Continued.

| XSA | popula number |  | (Thousand) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age |  |  |  |  |  |  |  |  |  |  |  |  |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |  |  |  |  |  |
| 2009 | 1576 | 3146 | 5730 | 29052 | 1239 | 419 | 786 |  |  |  |  |  |
| 2010 | 1606 | 1089 | 1936 | 3573 | 16506 | 566 | 598 |  |  |  |  |  |
| 2011 | 343 | 1177 | 409 | 622 | 1036 | 9807 | 940 |  |  |  |  |  |
| 2012 | 1370 | 278 | 910 | 267 | 344 | 684 | 8262 |  |  |  |  |  |
| 2013 | 36658 | 1117 | 221 | 615 | 166 | 279 | 1857 |  |  |  |  |  |
| 2014 | 38814 | 25584 | 827 | 90 | 426 | 100 | 1422 |  |  |  |  |  |
| 2015 | 20623 | 31411 | 18225 | 299 | 27 | 224 | 407 |  |  |  |  |  |
| 2016 | 10472 | 16812 | 23027 | 11856 | 151 | 16 | 83 |  |  |  |  |  |
| 2017 | 79118 | 8235 | 12813 | 16465 | 8813 | 121 | 121 |  |  |  |  |  |
| 2018 | 17509 | 63696 | 5232 | 8860 | 11498 | 6123 | 360 |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Estima | popula | ndance | at | 1st | Jan | 2019 |  |  |  |  |  |  |
| age |  |  |  |  |  |  |  |  |  |  |  |  |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |  |  |  |  |  |
| 2019 | 0 | 14256 | 46384 | 3827 | 5198 | 8873 | 4813 |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Fleet: | SCOGFS |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Log | catchal | duals. |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |  |
| 1 | -0.296 | 0.179 | -0.005 | -0.103 | 0.062 | 0.181 | -0.235 | NA | 0.121 | NA | -0.49 |  |
| 2 | -0.411 | 0.369 | 0.301 | -0.048 | 0.087 | 0.164 | -0.343 | NA | -0.265 | NA | -0.601 |  |
| 3 | -0.485 | 0.36 | 0.419 | 0.316 | 0.347 | -0.015 | -0.739 | NA | -0.276 | NA | -0.402 |  |
| 4 | -0.19 | 0.587 | 0.478 | 0.487 | 0.809 | -0.007 | -1.155 | NA | -0.339 | NA | -0.723 |  |
| 5 | -0.221 | 0.094 | 0.512 | -0.452 | 0.856 | -0.046 | -0.757 | NA | -0.439 | NA | -0.503 |  |
| 6 | 0.028 | 0.218 | -0.036 | -0.153 | 0.158 | -0.173 | -0.398 | NA | -0.193 | NA | -0.456 |  |
| year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 |  |
| 1 | -0.175 | 0.029 | NA | 0.383 | -0.15 | 0.26 | 0.264 | 0.308 | NA | 0.29 | -1.393 |  |
| 2 | -0.636 | 0.14 | NA | 0.165 | 0.632 | -0.287 | 0.219 | 0.408 | NA | 0.346 | 0.206 |  |
| 3 | -0.813 | -0.231 | NA | 0.019 | -0.045 | 0.422 | -0.003 | 0.309 | NA | 0.299 | 0.464 |  |
| 4 | -0.843 | -0.554 | NA | 0.553 | 0.572 | 0.681 | 0.25 | -0.046 | NA | 0.146 | -0.189 |  |
| 5 | -1.059 | 0.29 | NA | -0.474 | 0.868 | 0.112 | -0.179 | -0.458 | NA | 0.249 | 1.048 |  |
| 6 | -0.058 | 0.284 | NA | 0.089 | 0.285 | -0.145 | 0.015 | -0.444 | NA | 0.023 | -0.011 |  |
| year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |  |  |  |  |  |  |
| 1 | 0.371 | -0.051 | -0.057 | 0.129 | 0.142 | 0.235 |  |  |  |  |  |  |
| 2 | -1.363 | 0.326 | 0.264 | -0.067 | 0.299 | 0.095 |  |  |  |  |  |  |
| 3 | 0.331 | -1.06 | 0.69 | 0.096 | 0.021 | -0.025 |  |  |  |  |  |  |
| 4 | -1.66 | 0 | 0 | 0.582 | 0.203 | 0.36 |  |  |  |  |  |  |
| 5 | -0.137 | -0.612 | 0 | -0.037 | 0.092 | 0.13 |  |  |  |  |  |  |
| 6 | -0.383 | 0 | 0.181 | -0.095 | -0.081 | -0.247 |  |  |  |  |  |  |

Table 4.3.17. Continued.

|  | Mean |  | catchability and |  | standa error strengland |  | of | ages | with | catchability |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | indepe |  | year | class |  |  | constal | w.r.t. | time |  |  |
|  | 4 | 5 | 6 |  |  |  |  |  |  |  |  |
| Mean | -2.423 | -2.423 | -2.4228 |  |  |  |  |  |  |  |  |
| S.E_Lo\& | 0.625 | 0.521 | 0.2161 |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | Regres | statistic |  |  |  |  |  |  |  |  |  |
|  | Ages | with | q | depender | on | year | class | streng |  |  |  |
|  | slope | interce |  |  |  |  |  |  |  |  |  |
| Age | 1 | 0.701 | 4.162033 |  |  |  |  |  |  |  |  |
| Age | 2 | 0.773 | 3.847961 |  |  |  |  |  |  |  |  |
| Age | 3 | 0.847 | 3.495144 |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | Termin |  | survivor | and | F | summ | aries: |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | Age | 1 | Year | class | 2017 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| source |  |  |  |  |  |  |  |  |  |  |  |
|  | scaled | survive | yrcls |  |  |  |  |  |  |  |  |
| SCOGF | 0.698 | 19930 | 2017 |  |  |  |  |  |  |  |  |
| fshk | 0.215 | 4312 | 2017 |  |  |  |  |  |  |  |  |
| nshk | 0.088 | 18472 | 2017 |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | Age | 2 | Year | class | 2016 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| source |  |  |  |  |  |  |  |  |  |  |  |
|  | scaled | survive | yrcls |  |  |  |  |  |  |  |  |
| SCOGF | 0.712 | 52415 | 2016 |  |  |  |  |  |  |  |  |
| fshk | 0.288 | 37156 | 2016 |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | Age | 3 | Year | class | 2015 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| source |  |  |  |  |  |  |  |  |  |  |  |
|  | scaled | survivc | yrcls |  |  |  |  |  |  |  |  |
| SCOGF | 0.746 | 3716 | 2015 |  |  |  |  |  |  |  |  |
| fshk | 0.254 | 1126 | 2015 |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | Age |  | Year | class | 2014 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| source |  |  |  |  |  |  |  |  |  |  |  |
|  | scaled | survive | yrcls |  |  |  |  |  |  |  |  |
| SCOGF | 0.616 | 7448 | 2014 |  |  |  |  |  |  |  |  |
| fshk | 0.384 | 3711 | 2014 |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | Age | 5 | Year | class | 2013 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| source |  |  |  |  |  |  |  |  |  |  |  |
|  | scaled | survive | yrcls |  |  |  |  |  |  |  |  |
| SCOGF | 0.759 | 10102 | 2013 |  |  |  |  |  |  |  |  |
| fshk | 0.241 | 1989 | 2013 |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | Age |  | Year | class | 2012 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| source |  |  |  |  |  |  |  |  |  |  |  |
|  | scaled | survive | yrcls |  |  |  |  |  |  |  |  |
| SCOGF | 0.914 | 3760 | 2012 |  |  |  |  |  |  |  |  |
| fshk | 0.086 | 1090 | 2012 |  |  |  |  |  |  |  |  |

Table 4.3.18. Haddock in 6.b. Final XSA runs. Fishing mortality-at-age.

| table | $\begin{aligned} & \text { 4.3.18 } \\ & \hline 22 / 08 / 19 \\ & \hline \end{aligned}$ | 15:13:10 | HADDOCK LANDISC |  | 2014 | ROCKALL | . | harvest |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | units= | $f$ |  |  |  |  |  |
|  | age |  |  |  |  |  |  |  |  |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |  |  |
|  | 1991 | 0.24 | 0.61 | 0.901 | 0.948 | 0.46 | 0.695 | 0.695 |  |
|  | 1992 | 0.178 | 0.49 | 0.829 | 0.994 | 1.011 | 0.332 | 0.332 |  |
|  | 1993 | 0.107 | 0.337 | 0.662 | 0.543 | 0.986 | 0.82 | 0.82 |  |
|  | 1994 | 0.142 | 0.293 | 0.601 | 0.935 | 0.573 | 0.926 | 0.926 |  |
|  | 1995 | 0.051 | 0.253 | 0.572 | 0.797 | 0.836 | 0.362 | 0.362 |  |
|  | 1996 | 0.241 | 0.573 | 0.496 | 0.545 | 0.701 | 0.667 | 0.667 |  |
|  | 1997 | 0.167 | 0.342 | 0.316 | 0.349 | 0.617 | 1.081 | 1.081 |  |
|  | 1998 | 0.248 | 0.581 | 0.656 | 0.512 | 0.625 | 0.859 | 0.859 |  |
|  | 1999 | 0.501 | 0.754 | 0.898 | 0.948 | 0.881 | 1.434 | 1.434 |  |
|  | 2000 | 0.39 | 0.863 | 1.107 | 1.259 | 1.273 | 1.414 | 1.414 |  |
|  | 2001 | 0.115 | 0.154 | 0.256 | 0.21 | 1.057 | 1.31 | 1.31 |  |
|  | 2002 | 0.152 | 0.243 | 0.611 | 0.721 | 0.375 | 0.852 | 0.852 |  |
|  | 2003 | 0.167 | 0.438 | 0.579 | 0.632 | 1.134 | 0.854 | 0.854 |  |
|  | 2004 | 0.075 | 0.149 | 0.346 | 0.932 | 1.284 | 0.9 | 0.9 |  |
|  | 2005 | 0.084 | 0.482 | 0.438 | 0.509 | 0.167 | 0.289 | 0.289 |  |
|  | 2006 | 0.037 | 0.162 | 0.401 | 0.246 | 0.259 | 0.137 | 0.137 |  |
|  | 2007 | 0.133 | 0.17 | 0.75 | 0.823 | 0.378 | 0.34 | 0.34 |  |
|  | 2008 | 0.128 | 0.299 | 0.305 | 0.664 | 0.739 | 0.278 | 0.278 |  |
|  | 2009 | 0.17 | 0.286 | 0.272 | 0.365 | 0.583 | 1.182 | 1.182 |  |
|  | 2010 | 0.111 | 0.78 | 0.935 | 1.038 | 0.321 | 0.776 | 0.776 |  |
|  | 2011 | 0.012 | 0.057 | 0.227 | 0.394 | 0.215 | 0.343 | 0.343 |  |
|  | 2012 | 0.004 | 0.026 | 0.192 | 0.276 | 0.009 | 0.059 | 0.059 |  |
|  | 2013 | 0.16 | 0.101 | 0.698 | 0.168 | 0.3 | 0.405 | 0.405 |  |
|  | 2014 | 0.012 | 0.139 | 0.818 | 1.01 | 0.444 | 0.735 | 0.735 |  |
|  | 2015 | 0.004 | 0.11 | 0.23 | 0.485 | 0.328 | 0.36 | 0.36 |  |
|  | 2016 | 0.04 | 0.072 | 0.135 | 0.097 | 0.022 | 0.15 | 0.15 |  |
|  | 2017 | 0.017 | 0.254 | 0.169 | 0.159 | 0.164 | 0.763 | 0.763 |  |
|  | 2018 | 0.006 | 0.117 | 0.113 | 0.333 | 0.059 | 0.041 | 0.041 |  |

Table 4.3.19. Haddock in 6.b. Final XSA runs. Stock numbers ( ${ }^{*} 10^{3}$ ) at-age.

| table | 4.3.19 |  | HADDOCK LANDISC |  | 2014 | ROCKALL |  | stock.n |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 22/08/19 | 15:13:10 | units= | NA |  |  |  |  |  |
|  | age |  |  |  |  |  |  |  |  |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |  |  |
|  | 1991 | 109540 | 81935 | 28277 | 9640 | 5108 | 762 | 1136 |  |
|  | 1992 | 109143 | 70515 | 36457 | 9407 | 3059 | 2640 | 799 |  |
|  | 1993 | 121862 | 74805 | 35373 | 13033 | 2852 | 911 | 699 |  |
|  | 1994 | 68327 | 89657 | 43714 | 14941 | 6202 | 871 | 418 |  |
|  | 1995 | 61262 | 48549 | 54745 | 19623 | 4805 | 2862 | 285 |  |
|  | 1996 | 62439 | 47669 | 30854 | 25291 | 7238 | 1704 | 206 |  |
|  | 1997 | 71687 | 40176 | 22007 | 15388 | 12009 | 2941 | 339 |  |
|  | 1998 | 71695 | 49683 | 23361 | 13143 | 8890 | 5306 | 2626 |  |
|  | 1999 | 48472 | 45829 | 22755 | 9921 | 6450 | 3894 | 3930 |  |
|  | 2000 | 28158 | 24048 | 17661 | 7588 | 3148 | 2187 | 2944 |  |
|  | 2001 | 78115 | 15614 | 8305 | 4780 | 1763 | 722 | 844 |  |
|  | 2002 | 104614 | 57015 | 10962 | 5262 | 3171 | 502 | 842 |  |
|  | 2003 | 47320 | 73559 | 36619 | 4870 | 2095 | 1784 | 1140 |  |
|  | 2004 | 14170 | 32792 | 38866 | 16799 | 2119 | 552 | 869 |  |
|  | 2005 | 14506 | 10758 | 23128 | 22517 | 5418 | 480 | 845 |  |
|  | 2006 | 88242 | 10917 | 5440 | 12222 | 11084 | 3755 | 1297 |  |
|  | 2007 | 13174 | 69641 | 7603 | 2982 | 7822 | 7002 | 2739 |  |
|  | 2008 | 4368 | 9438 | 48124 | 2940 | 1072 | 4389 | 4208 |  |
|  | 2009 | 1576 | 3146 | 5730 | 29052 | 1239 | 419 | 786 |  |
|  | 2010 | 1606 | 1089 | 1936 | 3573 | 16506 | 566 | 598 |  |
|  | 2011 | 343 | 1177 | 409 | 622 | 1036 | 9807 | 940 |  |
|  | 2012 | 1370 | 278 | 910 | 267 | 344 | 684 | 8262 |  |
|  | 2013 | 36658 | 1117 | 221 | 615 | 166 | 279 | 1857 |  |
|  | 2014 | 38814 | 25584 | 827 | 90 | 426 | 100 | 1422 |  |
|  | 2015 | 20623 | 31411 | 18225 | 299 | 27 | 224 | 407 |  |
|  | 2016 | 10472 | 16812 | 23027 | 11856 | 151 | 16 | 83 |  |
|  | 2017 | 79118 | 8235 | 12813 | 16465 | 8813 | 121 | 121 |  |
|  | 2018 | 17509 | 63696 | 5232 | 8860 | 11498 | 6123 | 360 |  |

Table 4.3.20. Haddock in 6.b. Final XSA run. Summary table.

|  | RECRUITS | TOTALBIO | TOTSPBIO | LANDINGS | Yield/SSB | Fbar(2-5) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 109540 | 50576 | 15357 | 5656 | 0.3683 | 0.7295 |
| 1992 | 109143 | 49840 | 18471 | 5321 | 0.2881 | 0.8308 |
| 1993 | 121862 | 54043 | 19544 | 4781 | 0.2446 | 0.6318 |
| 1994 | 68327 | 55199 | 23854 | 5732 | 0.2403 | 0.6005 |
| 1995 | 61262 | 48513 | 28760 | 5587 | 0.1943 | 0.6148 |
| 1996 | 62439 | 44403 | 24438 | 7072 | 0.2894 | 0.5785 |
| 1997 | 71687 | 40589 | 21117 | 5167 | 0.2447 | 0.4058 |
| 1998 | 71695 | 43790 | 21795 | 4986 | 0.2288 | 0.5936 |
| 1999 | 48472 | 36338 | 18844 | 5356 | 0.2842 | 0.8703 |
| 2000 | 28158 | 22789 | 12730 | 5445 | 0.4278 | 1.1255 |
| 2001 | 78115 | 21243 | 5981 | 2020 | 0.3377 | 0.4194 |
| 2002 | 104614 | 36331 | 7155 | 3118 | 0.4357 | 0.4875 |
| 2003 | 47320 | 38443 | 14205 | 5968 | 0.4201 | 0.6958 |
| 2004 | 14170 | 27686 | 17971 | 6434 | 0.3581 | 0.6776 |
| 2005 | 14506 | 21446 | 17107 | 5239 | 0.3063 | 0.3988 |
| 2006 | 88242 | 26901 | 13346 | 2756 | 0.2065 | 0.2671 |
| 2007 | 13174 | 27314 | 12031 | 3347 | 0.2782 | 0.5302 |
| 2008 | 4368 | 22368 | 19862 | 4222 | 0.2126 | 0.5018 |
| 2009 | 1576 | 14298 | 13357 | 3241 | 0.2427 | 0.3766 |
| 2010 | 1606 | 12056 | 11548 | 3404 | 0.2947 | 0.7683 |
| 2011 | 343 | 8628 | 8274 | 1860 | 0.2248 | 0.2233 |
| 2012 | 1370 | 10547 | 10184 | 686 | 0.0674 | 0.1257 |
| 2013 | 36658 | 10975 | 2872 | 889 | 0.3096 | 0.3171 |
| 2014 | 38814 | 17434 | 2474 | 1845 | 0.7458 | 0.6028 |
| 2015 | 20623 | 22474 | 9371 | 2510 | 0.2679 | 0.2883 |
| 2016 | 10472 | 26020 | 18346 | 2504 | 0.1365 | 0.0815 |
| 2017 | 79118 | 43365 | 25077 | 4430 | 0.1767 | 0.1865 |
| 2018 | 17509 | 53527 | 28703 | 3850 | 0.1341 | 0.1556 |

Table 4.3.21. Haddock in 6.b. Detailed short-term forecast output.

| MFDP version 1a |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Run: FMSY1 |  |  |  |  |  |  |  |  |  |  |  |  |
| Time and date: 01:32 04.09.2019 |  |  |  |  |  |  |  |  |  |  |  |  |
| Fbar age range (Total) : 2-5 |  |  |  |  |  |  |  |  |  |  |  |  |
| Fbar age range Fleet 1: 2-5 |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Year: | 2018 | F multipli, | 1 | Fleet1 HCl | - 0.1085 | Fleet1 DFI | 0.047 |  |  |  |  |  |
|  | Catch |  |  |  |  |  |  |  |  |  |  |  |
| Age | F | CatchNos | Yield | DF | DCatchNo | DYield | StockNos | Biomass | SSNos(Jan | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.001 | 16 | 3 | 0.005 | 79 | 7 | 17509 | 2977 | 0 | 0 | 0 | 0 |
| 2 | 0.0397 | 2167 | 719 | 0.0773 | 4220 | 1089 | 63696 | 21848 | 0 | 0 | 0 | 0 |
| 3 | 0.0793 | 356 | 201 | 0.0337 | 151 | 55 | 5232 | 2742 | 5232 | 2742 | 5232 | 2742 |
| 4 | 0.2672 | 1835 | 1294 | 0.0658 | 452 | 191 | 8860 | 5316 | 8860 | 5316 | 8860 | 5316 |
| 5 | 0.0479 | 485 | 454 | 0.0111 | 112 | 54 | 11498 | 12912 | 11498 | 12912 | 11498 | 12912 |
| 6 | 0.0388 | 211 | 261 | 0.0022 | 12 | 6 | 6123 | 7109 | 6123 | 7109 | 6123 | 7109 |
| 7 | 0.0401 | 13 | 25 | 0.0009 | 0 | 0 | 360 | 623 | 360 | 623 | 360 | 623 |
| Total |  | 5083 | 2957 |  | 5027 | 1401 | 113278 | 53526 | 32073 | 28702 | 32073 | 28702 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Year: | 2019 | F multipli, | 1.1476 | Fleet1 HCl | - 0.1065 | Fleet1 DFI | 0.0555 |  |  |  |  |  |
|  | Catch |  |  |  |  |  |  |  |  |  |  |  |
| Age | F | CatchNos | Yield | DF | DCatchNo | DYield | StockNos | Biomass | SSNos(Jan | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.004 | 88 | 19 | 0.0201 | 440 | 65 | 24444 | 4620 | 0 | 0 | 0 | 0 |
| 2 | 0.0574 | 684 | 280 | 0.112 | 1334 | 343 | 14250 | 4902 | 0 | 0 | 0 | 0 |
| 3 | 0.1119 | 4360 | 2354 | 0.0476 | 1856 | 642 | 46392 | 22361 | 46392 | 22361 | 46392 | 22361 |
| 4 | 0.1807 | 563 | 345 | 0.0445 | 139 | 57 | 3826 | 2322 | 3826 | 2322 | 3826 | 2322 |
| 5 | 0.0761 | 343 | 386 | 0.0177 | 80 | 41 | 5199 | 5283 | 5199 | 5283 | 5199 | 5283 |
| 6 | 0.3454 | 2342 | 2721 | 0.0195 | 132 | 85 | 8874 | 10197 | 8874 | 10197 | 8874 | 10197 |
| 7 | 0.3571 | 1390 | 2406 | 0.0078 | 30 | 18 | 5095 | 8350 | 5095 | 8350 | 5095 | 8350 |
| Total |  | 9770 | 8512 |  | 4011 | 1251 | 108080 | 58034 | 69386 | 48513 | 69386 | 48513 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Year: | 2020 | F multipli | 1.4 | Fleet1 HCl | \| 0.13 | Fleet1 DFI | 0.0677 |  |  |  |  |  |
|  | Catch |  |  |  |  |  |  |  |  |  |  |  |
| Age | F | CatchNos | Yield | DF | DCatchNo | DYield | StockNos | Biomass | SSNos(Jan | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0049 | 62 | 14 | 0.0245 | 310 | 46 | 14170 | 2678 | 0 | 0 | 0 | 0 |
| 2 | 0.07 | 1124 | 460 | 0.1366 | 2193 | 564 | 19537 | 6721 | 0 | 0 | 0 | 0 |
| 3 | 0.1365 | 1111 | 600 | 0.0581 | 473 | 164 | 9849 | 4747 | 9849 | 4747 | 9849 | 4747 |
| 4 | 0.2205 | 5684 | 3479 | 0.0543 | 1400 | 574 | 32382 | 19656 | 32382 | 19656 | 32382 | 19656 |
| 5 | 0.0928 | 199 | 224 | 0.0216 | 46 | 24 | 2501 | 2541 | 2501 | 2541 | 2501 | 2541 |
| 6 | 0.4214 | 1204 | 1399 | 0.0238 | 68 | 44 | 3876 | 4453 | 3876 | 4453 | 3876 | 4453 |
| 7 | 0.4357 | 2549 | 4413 | 0.0095 | 56 | 34 | 7940 | 13014 | 7940 | 13014 | 7940 | 13014 |
| Total |  | 11933 | 10587 |  | 4547 | 1448 | 90254 | 53809 | 56547 | 44411 | 56547 | 44411 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Year: | 2021 | F multipli | 1.4 | Fleet1 HCl | \| 0.13 | Fleet1 DFI | 0.0677 |  |  |  |  |  |
|  | Catch |  |  |  |  |  |  |  |  |  |  |  |
| Age | F | CatchNos | Yield | DF | DCatchNo | DYield | StockNos | Biomass | SSNos(Jan | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0049 | 62 | 14 | 0.0245 | 310 | 46 | 14170 | 2678 | 0 | 0 | 0 | 0 |
| 2 | 0.07 | 648 | 265 | 0.1366 | 1265 | 325 | 11265 | 3875 | 0 | 0 | 0 | 0 |
| 3 | 0.1365 | 1467 | 792 | 0.0581 | 625 | 216 | 13009 | 6270 | 13009 | 6270 | 13009 | 6270 |
| 4 | 0.2205 | 1165 | 713 | 0.0543 | 287 | 118 | 6638 | 4029 | 6638 | 4029 | 6638 | 4029 |
| 5 | 0.0928 | 1604 | 1808 | 0.0216 | 373 | 193 | 20142 | 20464 | 20142 | 20464 | 20142 | 20464 |
| 6 | 0.4214 | 567 | 659 | 0.0238 | 32 | 21 | 1826 | 2098 | 1826 | 2098 | 1826 | 2098 |
| 7 | 0.4357 | 1990 | 3445 | 0.0095 | 43 | 26 | 6198 | 10159 | 6198 | 10159 | 6198 | 10159 |
| Total |  | 7503 | 7695 |  | 2935 | 944 | 73248 | 49573 | 47812 | 43020 | 47812 | 43020 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Input units are thousands and kg - output in tonnes |  |  |  |  |  |  |  |  |  |  |  |  |

Table 4.3.22. Haddock in 6.b. Input data for the short-term forecast.

| MFDP version 1a |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Run: FMSY1 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| Time and date: 01:32 04.09.2018Fbar age range (Total) : 2-5 |  |  |  |  |  |  |
| Fbar age range Fleet 1:2-5 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| 2018 |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | sWt |
| 1 | 17509 | 0.2 | 0 | 0 | 0 | 0.17 |
| 2 | 63696 | - 0.2 | 0 | 0 | 0 | 0.0343 |
| 3 | 5232 | 0.2 | 1 | 0 | 0 | 0.524 |
| 4 | 8860 | 0.2 | 1 | 0 | 0 | 0.6 |
| 5 | 11498 | 0.2 | 1 | 0 | 0 | - 1.123 |
| 6 | 6123 | 0.2 | 1 | 0 | 0 | - 1.161 |
| 7 | 360 | - 0.2 | 1 | 0 | 0 | - 1.731 |
|  |  |  |  |  |  |  |
| Catch |  |  |  |  |  |  |
| Age | Sel | cWt | DSel | DCWt |  |  |
| 1 | 0.000989 | 0.219 | 0.005 | 0.087 |  |  |
| 2 | 0.0397 | 0.332 | 0.0773 | 0.258 |  |  |
| 3 | 0.0793 | 0.564 | 0.0337 | 0.361 |  |  |
| 4 | 0.2672 | - 0.705 | 0.0658 | 0.422 |  |  |
| 5 | 0.0479 | 0.935 | 0.0111 | 0.479 |  |  |
| 6 | 0.0388 | 1.235 | 0.0022 | 0.536 |  |  |
| 7 | 0.0401 | 1.928 | 0.0009 | 0.603 |  |  |
|  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| 2019 |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | sWt |
| 1 | 24444 | 0.2 | 0 | 0 | 0 | - 0.189 |
| 2. |  | 0.2 | 0 | 0 | 0 | - 0.344 |
| 3. |  | 0.2 | 1 | 0 | 0 | - 0.482 |
| 4. | . | 0.2 | 1 | 0 | 0 | - 0.607 |
| 5. | . | 0.2 | 1 | 0 | 0 | - 1.016 |
| 6. | . | 0.2 | 1 | 0 | 0 | - 1.149 |
| 7. | . | 0.2 | 1 | 0 | 0 | - 1.639 |
|  |  |  |  |  |  |  |
| Catch |  |  |  |  |  |  |
| Age | Sel | cWt | DSel | DCWt |  |  |
| 1 | 0.0035 | 0.219 | 0.0175 | 0.147 |  |  |
| 2 | 0.05 | 0.409 | 0.0976 | 0.257 |  |  |
| 3 | 0.0975 | 0.54 | 0.0415 | 0.346 |  |  |
| 4 | 0.1575 | 0.612 | 0.0388 | 0.41 |  |  |
| 5 | 0.0663 | 1.127 | 0.0154 | 0.517 |  |  |
| 6 | 0.301 | 1.162 | 0.017 | 0.642 |  |  |
| 7 | 0.3112 | 1.731 | 0.0068 | 0.603 |  |  |
|  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| 2020 |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | swt |
| 1 | 14170 | 0.2 | 0 | 0 | 0 | - 0.189 |
| 2. |  | 0.2 | 0 | 0 | 0 | - 0.344 |
| 3. |  | 0.2 | 1 | 0 | 0 | - 0.482 |
| 4. |  | 0.2 | 1 | 0 | 0 | - 0.607 |
| 5. |  | 0.2 | 1 | 0 | 0 | - 1.016 |
| 6. | . | 0.2 | 1 | 0 | 0 | - 1.149 |
| 7. | . | 0.2 | 1 | 0 | 0 | - 1.639 |
|  |  |  |  |  |  |  |
| Catch |  |  |  |  |  |  |
| Age | Sel | cWt | DSel | DCWt |  |  |
| 1 | 0.0035 | 0.219 | 0.0175 | 0.147 |  |  |
| 2 |  | 0.409 | 0.0976 | 0.257 |  |  |
| 3 | 0.0975 | 0.54 | 0.0415 | 0.346 |  |  |
| 4 | 0.1575 | 0.612 | 0.0388 | 0.41 |  |  |
| 5 | 0.0663 | 1.127 | 0.0154 | 0.517 |  |  |
| 6 | 0.301 | 1.162 | 0.017 | 0.642 |  |  |
| 7 | 0.3112 | 1.731 | 0.0068 | 0.603 |  |  |
|  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| 2021 |  |  |  |  |  |  |
| Age | $N$ | M | Mat | PF | PM | swt |
| 1 | 14170 | 0.2 | 0 | 0 | 0 | 0 0.189 |
| 2. | 2. | 0.2 | 0 | 0 | 0 | - 0.344 |
| 3. | . | 0.2 | 1 | 0 | 0 | - 0.482 |
| 4. |  | 0.2 | 1 | 0 | 0 | - 0.607 |
| 5. |  | 0.2 | 1 | 0 | 0 | - 1.016 |
| 6. | . | 0.2 | 1 | 0 | 0 | - 1.149 |
| 7. | 7. | 0.2 | 1 | 0 | 0 | - 1.639 |
|  |  |  |  |  |  |  |
| Catch |  |  |  |  |  |  |
| Age | Sel | cWt | DSel | DCWt |  |  |
| 1 | 0.0035 | 0.219 | 0.0175 | 0.147 |  |  |
| 2 |  | 0.409 | 0.0976 | 0.257 |  |  |
| 3 | 0.0975 | 0.54 | 0.0415 | 0.346 |  |  |
| 4 | 0.1575 | 0.612 | 0.0388 | 0.41 |  |  |
| 5 | 0.0663 | 1.127 | 0.0154 | 0.517 |  |  |
| 6 | 0.301 | 1.162 | 0.017 | 0.642 |  |  |
| 7 | 0.3112 | 1.731 | 0.0068 | 0.603 |  |  |
|  |  |  |  |  |  |  |

Table 4.3.23. Haddock in 6.b. Short-term forecast output.

| MFDP version 1a |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Run: FMSY |  |  |  |  |  |  |  |  |
| Time and date: 08:19 26.08.2019 |  |  |  |  |  |  |  |  |
| Fbar age range (Total) : 2-5 |  |  |  |  |  |  |  |  |
| Fbar age range Fleet 1:2-5 |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |
| 2018 |  |  |  |  |  |  |  |  |
|  |  | Catch | Landings |  | Discards |  |  |  |
| Biomass | SSB | FMult | FBar | Yield | FBar | Yield |  |  |
| 53526 | 28702 | 1 | 0.1085 | 2957 | 0.047 | 1401 |  |  |
| 2019 |  |  |  |  |  |  |  |  |
|  |  | Catch | Landings |  | Discards |  |  |  |
| Biomass | SSB | FMult | FBar | Yield | FBar | Yield |  |  |
| 58034 | 48513 | 1.1476 | 0.1065 | 8512 | 0.0555 | 1251 |  |  |
|  |  |  |  |  |  |  |  |  |
| 2020 |  |  |  |  |  |  | 2021 |  |
|  |  | Catch | Landings |  | Discards |  |  |  |
| Biomass | SSB | FMult | FBar | Yield | FBar | Yield | Biomass | SSB |
| 53809 | 44411 | 1.1 | 0.1021 | 8614 | 0.0532 | 1165 | 52364 | 45786 |
| . | 44411 | 1.13 | 0.1049 | 8818 | 0.0546 | 1194 | 52076 | 45500 |
| . | 44411 | 1.16 | 0.1077 | 9020 | 0.0561 | 1223 | 51790 | 45217 |
| . | 44411 | 1.19 | 0.1105 | 9221 | 0.0575 | 1251 | 51506 | 44936 |
| . | 44411 | 1.22 | 0.1132 | 9420 | 0.059 | 1280 | 51224 | 44656 |
| . | 44411 | 1.25 | 0.116 | 9618 | 0.0604 | 1308 | 50945 | 44379 |
| . | 44411 | 1.28 | 0.1188 | 9815 | 0.0619 | 1337 | 50667 | 44103 |
| . | 44411 | 1.31 | 0.1216 | 10010 | 0.0633 | 1365 | 50390 | 43830 |
| - | 44411 | 1.34 | 0.1244 | 10204 | 0.0648 | 1393 | 50116 | 43558 |
| - | 44411 | 1.37 | 0.1272 | 10396 | 0.0662 | 1420 | 49844 | 43288 |
| . | 44411 | 1.4 | 0.13 | 10587 | 0.0677 | 1448 | 49573 | 43020 |
| . | 44411 | 1.43 | 0.1327 | 10777 | 0.0691 | 1476 | 49305 | 42754 |
| . | 44411 | 1.46 | 0.1355 | 10965 | 0.0706 | 1503 | 49038 | 42489 |
| . | 44411 | 1.49 | 0.1383 | 11152 | 0.072 | 1530 | 48773 | 42227 |
| - | 44411 | 1.52 | 0.1411 | 11338 | 0.0735 | 1558 | 48510 | 41966 |
| . | 44411 | 1.55 | 0.1439 | 11522 | 0.0749 | 1585 | 48248 | 41707 |
| . | 44411 | 1.58 | 0.1467 | 11705 | 0.0764 | 1612 | 47989 | 41450 |
| - | 44411 | 1.61 | 0.1494 | 11886 | 0.0778 | 1638 | 47731 | 41194 |
| . | 44411 | 1.64 | 0.1522 | 12067 | 0.0793 | 1665 | 47475 | 40941 |
| - | 44411 | 1.67 | 0.155 | 12246 | 0.0807 | 1691 | 47220 | 40689 |
|  | 44411 | 1.7 | 0.1578 | 12423 | 0.0822 | 1718 | 46967 | 40438 |
| . |  |  |  |  |  |  |  |  |
| Input units are thousands and kg - output in tonnes |  |  |  |  |  |  |  |  |

Table 4.3.24. Haddock 6.b. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.



Figure 4.3.1. Distribution of haddock (catch N per 30 minutes) on the Rockall Bank in 2001-2017 from the Scottish trawl survey (Scottish Rock-IBTS-Q3).


Figure 4.3.2. Haulings pattern during bottom survey by RV 'Scotia' in September 2017.


Figure 4.3.3. Distribution of 0-group haddock (number per 30 minutes) on the Rockall Bank in 2012-2018 from the Scottish trawl survey.


Figure 4.3.4. Abundance (a) and biomass (b) of haddock, assessed with the trawl survey method with geographical stratification based on rectangles of 15 ' latitude and 15 ' longitude by RV 'Scotia' survey. Red dashed line indicates the confidence interval with 0.95 reliability level.


Figure 4.3.5. Abundance (a) and biomass (b) of haddock, assessed with the trawl survey method with geographical stratification based on bathymetry by RV 'Scotia' survey. Red dashed line indicates the confidence interval with 0.95 reliability level.


Figure 4.3.6. Abundance (a) and biomass (b) of haddock, assessed with the trawl survey method without geographical stratification by RV 'Scotia' survey. Red dashed line indicates the confidence interval with 0.95 reliability level.


Figure 4.3.7. Rockall haddock in 6.b. Scottish, Irish effort in 1985-2016 and Russian effort in 1999-2016. Data for 2017-2018 are not available.


Figure 4.3.8. Lpue and cpue of the fleets fishing for Rockall haddock. Note that Scottish and Irish effort data are not reliable because reporting is not mandatory (data for 2016-2018 are not available).
1 - Scottish lpue (all gears).
2 - Irish trawlers lpue.
3 - Cpue of Russian trawlers (BMRT type, tonnage class 10 in 1999-2007, and tonnage class 9 in 2008-2009,2013-2015).


Figure 4.3.9. Dynamics of haddock total biomass (ICES, 2008a; ICES, 2008b) and directed fishing efficiency ( $t$ per a trawling hour) for tonnage class 10 vessels in 1999-2007.


Figure 4.3.10. Total landings and discards of Rockall haddock ('000 individuals).


Figure 4.3.11. Total landings and discards of Rockall haddock (tonnes).


Figure 4.3.12. Comparison the landings, discards and catch of Rockall haddock from assessments 2018 and 2019.


Figure 4.3.13. Haddock in 6.b. Mean weights-at-age in discards (left) and in landings (right).


Figure 4.3.14. Haddock in 6.b. Mean weights-at-age in catch (left) and in stock (right)


Figure 4.3.15. Haddock in 6.b. Log catch (with discards in numbers)-at-age by year.


Figure 4.3.16. Haddock in 6.b. Log landings (in numbers) at age by year.


Figure 4.3.18. Haddock in 6.b. Catch curves (with registered discards).


Figure 4.3.19. Haddock in 6.b. Catch curves (with registered discards).


Figure 4.3.20. Haddock in 6.b. Relative catch proportion-at-age.


Figure 4.3.21. Haddock in 6.b. Log survey cpue at-age by year.


Figure 4.3.22. Haddock in 6.b. Final XSA run. Log survey cpue by year class.


Figure 4.3.23. Haddock in 6.b. Log survey cpue-at-age.

## SCOGFS



Figure 4.3.24. Haddock in 6.b. The analy sis of survey data. Pairwise plots of age.


Figure 4.3.25. Haddock in 6.b. Log catchability residual plots (shrinkage 1.0, catchability dependent on stock size at-ages $<4$ ). Final XSA.


Figure 4.3.26. Haddock in 6.b. Adjusted Scottish groundfish survey cpue from the final XSA run plotted against VPA numbers(shrinkage 1.0) at-age. Catchability dependent on stock size at-ages $<4$.


Figure 4.3.27. Haddock in 6.b. Survey indices and XSA estimates (shrinkage 1.0) at-age. Final XSA: catchability dependent on stock size at-ages $<4$.


Figure 4.3.28. Haddock in 6.b. Retrospective analy ses (F shrinkage 1.0).


Figure 4.3.29. Haddock in 6.b. XSA assessment. Summary plots.


Figure 4.3.30. Haddock in 6.b. Comparison of the current final assessment (in red) with the previous one (in black). In the SSB plot, the solid blue line indicates $B_{P A}$ and the dotted blue line refers to $B_{l i m}$. In the fishing mortality plot, the solid blue line signifies Fra.


Figure 4.3.31. Haddock in 6.b. Comparison of observed and predicted survey and catch biomass derived from StatCam. The parametric model (scenario 2).


Figure 4.3.32. Haddock in 6.b. The SSB and recruitment by StatCam estimation. The parametric model (scenario 2).


Figure 4.3.33. Haddock in 6.b. The SSB, recruitment and fishing mortality by SAM estimation.


Figure 4.3.34. Haddock in 6.b. Comparison of the final XSA (VPA) assessment with the statistical catch-at-age model StatCam and SAM assessments.


Figure 4.3.35. Haddock in 6.b. Scottish Groundfish survey indices of haddock abundance-at-age 0 .


Figure 4.3.36. Haddock in 6.b. VPA numbers-at-age 1 from XSA plotted against Scottish Groundfish survey indices of haddock at-age 0 .


Figure 4.3.37. Haddock in Division 6.b. Discard proportion-at-age by year, and mean discard proportion-at-age for periods: 1991-2018, 1999-2018, 2006-2018 and 2009-2018.

# 8 Haddock (Melanogrammus aeglefinus) in Division 7.a (Irish Sea) 

## Type of assessment

Age-structured assessment model using Age Structured Assessment Program (ASAP).

## ICES advice applicable to 2018

ICES advises that when the MSY approach is applied, catches in 2018 should be no more than 3444 tonnes.

ICES advice applicable to 2019
ICES advises that when the MSY approach is applied, catches in 2019 should be no more than 3739 tonnes.

### 8.1 General

## Stock descriptions and management units

The stock and management units are both ICES Division 7.a (Irish Sea). Landing taken or reported by Irish vessels in the southern most rectangles of 7 .a have been reassigned to the $7 . \mathrm{b}-\mathrm{k}$ stock since 2003 because they are believed to be part of the Celtic Sea stock.

## Management applicable to 2019

Management measures include TAC and effort restrictions as well as technical measures. Due to the bycatch of cod in the haddock fishery, the regulations affecting Irish Sea haddock remain linked to those implemented under the cod recovery plan. From 1st January 2019, all fleets catching haddock are subject to the landing obligation.

TAC regulations for 2018 and 2019 are given below:

2018 management (Council Regulation (EU) 2018/127)

| Species:Haddock <br> Melanogrammus acglefinus Zone: 7a <br> (HAD/07A.) <br> Belgium 51  <br> France 232  <br> Ireland 1388  <br> United Kingdom 1536  <br> Union 3207  <br> TAC 3207 Analytical TAC <br>   Article 7(2) of this Regulation applies |
| :--- | ---: | :--- | :--- |

2019 management (Council Regulation (EU) 2019/124)

| Species:Haddock <br> Melanogrammus aeglefinus Zone:7a <br> (HAD/07A.)  <br> Belgium 59  <br> France 271  <br> Ireland 1619  <br> United Kingdom 1790  <br> Union 3739  <br> TAC 3739 Analytical TAC <br>   Article 7(2) of this Regulation applies |
| :--- | :--- | :--- | :--- |

The minimum landing size for haddock in the Irish Sea is 30 cm .

## Landings obligation

Since 2017 the landings obligation has applied to the stock. According to the delegate regulation (EC, 2015) vessels where more than $25 \%$ of their landings using trawls and seines in the reference years (2013 and 2014) and area were specified gadoids (cod, haddock, whiting and saithe) were covered by the Landings Obligation. This implies that all catches of haddock in the Irish Sea by those vessels must be landed. From the 1st January 2019, all fleets catching haddock will be subject to the landings obligation.

## Fishery in 2018

The characteristics of the fishery are described in the stock annex.
The fishery in 2018 was prosecuted by a similar fleet and gears as in recent years, with directed fishing prevented inside the cod closure in spring. The targeted whitefish fishery that developed during the 1990 using semi-pelagic trawls was in decline underwent but since 2014 there has been a slight increase in activity due to abundance of the haddock stock and increased fishing opportunity. However, this continues to be pursued by a small number of vessel (<15). A proportion of the TAC is taken as bycatch in the Nephrops fishery in a mixed fishery.

In 2018 , the uptake of TAC was $79 \%$. The primary two nations exploiting the stock are the UK and Ireland. The UK used $102 \%$ of quota allocation whilst Ireland used $69 \%$.ICES catch estimates are adjusted for reallocation of Irish landings from southern rectangles of 7.a to 7.g, as it is believed that these fish do not belong to the 7 .a stock. Table 11.1 gives nominallandings of haddock from the Irish Sea (Division 7.a) as reported by each country to ICES since 1984.

### 8.2 Data

All required to perform the assessment were supplied as required. Data submitted to InterCatch was used for allocation and raising of unsampled fleets. The unsampled fleets for discard estimates $19.5 \%$ of the total discard estimate. Age sampling was carried out on those fleets contributing $>95 \%$ of the international landings and $>80 \%$ of the international discards total in 2018 . The assessment uses landings and discard information, updated proportion mature-at-age estimates and updated tuning indices.

## Landings

Table11.2 gives the long-term trend of nominal landings of haddock from the Irish Sea (Division 7.a) as reported to ICES since 1972, together with Working Group estimates. The 1993-2005 WG estimates includes sampled-based re-estimates of landings into the main IrishSea ports. Samplebased evidence suggests that WG estimates are similar to reported landings since 2006. Following the benchmark (WKROUND 2013) the landings have been revised since 1993, and exclude landings from the southern rectangles in the Irish Sea, as they not are believed to be part of this stock.

The methods for estimating quantities and composition of haddock landings from 7.a, used in previous years, are described in the stock annex (see Annex 2). The series of numbers-at-age in the international commercial catch is given in Table 11.3. Sampling levels were not considered adequate to derive catch age compositions in 2003. The time-series mean weight-at-age in the catch is given Table 11.4.

## Discards

Annual discard data were updated for Ireland and Northern Ireland. Historic discard numbers-at-age for the different sampled fleets are given in the stock annex (see Annex 2). Issues relating to the reliability and confidence in the data were addressed at the benchmark assessment for this stock (WKROUND 2013; WKIrish3 2017).

Methods for estimating quantities and composition of discards from UK (NI) and Irish Nephrops trawlers are described in the stock annex. Sampling levels have increased in recent years. The large estimates of discarding for Nephrops fleets observed by previous WG are still evident. A historic time-series of discard numbers-at-age was constructed at the benchmark. Discard rates are very variable between fleets.

## Biological data

The derivation of biological parameters and variables is described in the stock annex (see Annex 2). Natural mortality-at-age was calculated using the methods proposed by Lorenzen (1996) at WKIrish2 (2016). The proportions mature-at-age was also recalculated at the benchmark, and based on the mean proportion observed during the NIGFS-WIBTS-Q1 survey with a smoother fitted this is updated annually.

There is evidence of trends in mean length-at-age over time (Figure 11.1), which needs to be reflected in the stock weights-at-age. Since 2001, the WG calculated stock weights by fitting a von Bertalanffy growth curve to survey estimates of mean length-at-age in March, described in the stock annex. The procedure was updated this year using NIGFS-WIBTS-Q1 (2018) and quarter one commercial landings data for 2018. The time-series of length-weight parameters indicate a reduction in expected weight-at-length since 1996, although this strength of this decline has reduced in recent years (see stock annex for historical data):

|  | Length-weight parameters |  | Expected weight-at-length |  |
| :--- | :--- | :--- | :--- | :--- |
| Year | A |  |  |  |

The following parameter estimates were obtained:
Mean $\mathrm{LI}_{y c}=45.4 \mathrm{~cm} ; \mathrm{K}=0.428 ; \mathrm{t}_{0}=-0.092$

Year-class effects giving estimates of asymptotic length relative to the mean were as follows:

| Year class | Effect | Year class | Effect |
| :---: | :---: | :---: | :---: |
| 1990 | 0.949 | 2004 | 0.983 |
| 1991 | 0.979 | 2005 | 0.989 |
| 1992 | 0.954 | 2006 | 0.953 |
| 1993 | 1.045 | 2007 | 0.986 |
| 1994 | 1.092 | 2008 | 0.961 |
| 1995 | 1.018 | 2009 | 1.002 |
| 1996 | 0.968 | 2010 | 1.058 |
| 1998 | 1.024 | 2011 | 1.074 |
| 2099 | 0.004 | 2012 | 1.106 |
| 2001 | 0.971 | 2013 | 1.014 |
| 2002 | 0.971 | 2015 | 1.019 |
|  |  | 2017 | 0.943 |
|  |  |  | 0.920 |

The year-class effects show a smooth decline from the mid-1990s coinciding with the rapid growth of the stock and may represent density-dependent growth effects, although other environmental factors may contribute. There is evidence in a reversal of this trend in recent years.

The resultant stock weights-at-age are given in Table 11.3. The weight-at-age in the stock shows a decreasing trend over time, which appears to have reversed in recent years.

## Surveys

The survey data considered in the assessment for this stock are given in Table 11.5. All survey series data for haddock available to the Working Group are described in the stock annex (see Annex 2). The following age-structured abundance indices were used in the assessment:

- UK (NI) groundfish survey (NIGFS) in March (age classes 1 to 4, years 1992-2018). Acronym NIGFS-WIBTS-Q1.
- UK (NI) groundfish survey (NIGFS) in October (age classes 0 to 3; years 1991 to 2018). Acronym NIGFS-WIBTS-Q4.
- UK (NI) Methot-Isaacs-Kidd (NI-MIK) net survey in June (age 0; years 1994-2018).
- UK Fishery Science Partnership (UKFspW) western Irish Sea roundfish survey (age classes 2 to 5, years 2004-2018, the survey was not conducted in 2014).

The relative log standardised indices for cohorts are plotted against time in Figure 11.2. Whilst ages 2 to 4 appear to show strong signal in the UKFspW the ability to detect the year class in age 5 haddock is less clear. The strong2013 year class continues to be tracked in all indices, indicating that the different surveys are capturing the prominent year-class signals in this stock (Figure 11.2). Correlation between survey indices by age is positive for all surveys and show high consistency within each survey (Figure 11.3). The indices from the UKFspW survey in the western IrishSea also show similar year-class signals to the other survey-series, but are noisy with strong year effects (Figure 11.2).

### 8.3 Assessment

## Deviation from stock annex

The assessment presented is the single fleet ASAP model. Recent changes in targeted fishing activity were accounted for with the inclusion of an additional selectivity pattern (2013-2017) in the 2018 assessment. In 2019, there was an apparent poor fit of the model to tuning series age structure and catch-at-age data, in particular for age 0 haddock. Figure 11.4A shows the residuals of the model-fit to catch-at-age data. A pattern of large negative residuals was observed for age ' 0 ' haddock 2013-2018. It was considered that the selectivity block introduced to the model in 2018 (2013-present) may be fitted to over select age 0 fish in the catch-at-age data, resulting in here a biased prediction this age class compared to the observed catch data.

The assessment model estimated selectivity of age 0 haddock in the 2019 assessment was compared to that estimated in 2018. In 2019, it was estimated as 0.19 compared to 0.08 in 2018. Within the model setup, the initial starting value of the selectivity pattern is set as 0.3 . The model fit when provided with a lower starting value of 0.1 was compared to the update assessment. The lower starting value improved the model fit with a reduced objective function from 1673 to 1654. The estimated selectivity of age 0 haddock was re-estimated by the model as 0.08 in the 20132018 selectivity block. The observed residual pattern was improved (Figure11.4) with overall reduction in the residuals and reduction of the bias in the final selectivity block.

ASAP was used for the assessment and model settings:

| Option | Setting |
| :---: | :---: |
| Use likelihood constant | Yes |
| Mean $F\left(F_{\text {bar }}\right)$ age range | 2-4 |
| Fleet selectivity block 1 | Asymptotic |
| Fleet selectivity block 2 | Age coefficients (age 0-5) (0.2;0.5;0.8;1;0.7;0.5) |
| Fleet selectivity block 3 | Age coefficients (age 0-5) (0.3;0.6;0.7;0.7;0.4;0.2) |
| *Fleet selectivity block 4 | Age coefficients (age 0-5) (0.1;0.6;0.8;0.9;1.0;1.0) |
| Discards | Included in catch (not specified separately from landings) |
| Index units | 4 (numbers) |
| Index month | NIGFS-Q1 (3); NIGFS-Q4 (10); NIMIK (7); UKFSPW(3) |
| Index selectivity linked to fleet | -1 (not linked) |
| Index age range | NIGFS-Q1 (1-4); NIGFS-Q4 (0-3); NIMIK (0); UKFSPW(2-5) |
| Index Selectivity (NIGFS-Q1) | Double logistic |
| Index Selectivity (NIGFS-Q4) | Asymptotic |
| Index Selectivity (NIMIK) | NA (age 0 only) |
| Index Selectivity (UK-FSPW) | Asymptotic |
| Index CV \& ESS (NIGFS-Q1) | Observed strata CV (lower limit 0.1); ESS = 50 |
| Index CV \& ESS (NIGFS-Q4) | Observed strata CV (lower limit 0.1); ESS = 50 |
| Index CV \& ESS (NIMIK) | Observed station CV (lower limit 0.1); ESS = 50 |
| Index CV \& ESS (UK-FSPW) | $C V=0.7 ; E S S=10$ |
| Phase for F-Mult in 1st year | 1 |
| Phase for F-Mult deviations | 2 |
| Phase for recruitment deviations | 3 |
| Phase for N in 1st Year | 1 |
| Phase for catchability in 1st Year | 3 |
| Phase for catchability deviations | -5 (Assume constant catchability in indices) |
| Phase for unexploited stock size | 1 |
| Phase for steepness | -5 (Do not fit stock-recruitment curve) |
| Catch total CV | 1993-2000 (0.175); 2003-2006 (0.2); 2007-present (0.15) |


| Option | Setting |
| :--- | :--- |
| Catch effective sample size | $1993-2000$ (50); 2003-2006 (1); 2007-present (50) |
| Lambda for recruit deviations | 0 (freely estimated) |
| Lambda for total catch | 1 |
| Lambda for total discards | NA (discards included in catch) |
| Lambda for F-Mult in 1st year | 0 (freely estimated) |
| Lambda for F-Mult deviations | 0 (freely estimated) |
| Lambda for index | 1 for both indices in the model |
| Lambda for index catchability | 0 for all indices (freely estimated) |
| Lambda for catchability devs | NA (phase is negative) |
| Lambda N in 1st year deviations | 0 (freely estimated) |
| Lambda devs initial steepness | 0 (freely estimated) |
| Lambda devs unexpl stock size | 0 (freely estimated) |

*Changed in 2019 compared to 2018.

## Final update assessment

The final assessment was run with the same settings as established by WKIrish 2017 and described in the stock annex, with the addition of a new selectivity pattern 2013-2017, as applied in 2018 and with a the lower starting value for selection of age 0 haddock in the final selectivity block. Discards were combined with the landings as catch in the model.

Figure 11.5 shows the predicted and observed catch. The catch information from 2007 to present is regarded as the most confident, during 2003-2006 it is regarded that catch and sampling information is of relatively lower quality due to lack of sampling opportunity. Before 2003, the catch series is regarded as of intermediate confidence. The model has close fit to the current observed catch 2011-present. Before this time, there is consistent over estimation of the catch 2000-2011 following a period of consistent underestimation of catch 1993-2000. Figure 11.6 shows the residuals of the catch proportions-at-age. For all ages there appears to good fit with no consistent pattern, however, there are some large deviations from observed and predicted for age 5 fish across the series. Figure 11.7 shows that the catch is dominated by fish $<4$ years, therefore the large residuals for fish of age 5 are likely to result from low sampling and small contribution of $5+$ fish to the stock. The fishing pressure ( F )-at-age is shown in Table 11.6.
The residuals of the index are shown in Figure 11.7. A good fit to the NI-MIK index is seen across the series, although some single year events are observed with a strong deviation in the last two years of the index. For the UKFsPW survey a poor fit to the 2009, 2017 and 2018 is evident. This suggests an inability of the model to track the large survey index values, this should be investigated further to explore the method of index calculation. During the most recent two years of the index, when the stock biomass has been high the UKFspW survey appears to tend of overestimated compared to the model fit. There is strong tracking of the both NIGFS-WIBTS-Q1 and NIGFS-WIBTS-Q4 index in general patterns, however, a general trend to under estimate the NIGFS-WIBTS-Q4 index by the model is observed whilst the NIGFS-WIBTS-Q1 shows an initial period of over estimation 1993-2000, followed by a period of under estimation 2002-2013.

Figure 11.9 shows the residuals of the survey proportions-at-age. For all indices, there is close fit between the observed and model predicted fit for fish up to four years old. The largest deviations occur in five year old fish in the UKFspW survey, which under reported five year old fish prior to 2014.

Figure 11.10 shows the retrospective analysis. The predicted catch shows no obvious retrospective pattern, neither does the recruitment estimate or fishing pressure. However, the SSB has a tendency to be revised downwards. The historic widely splayed retrospective runs are caused by re-estimation of selectivity patterns with a short terminal selectivity blocks (four years) and the influence of decreasing UK-FspW tuning series length with each retro peel. The results of the assessment are given in Table 11.8.

## Comparison with previous assessments

Figure 11.11 shows the comparison of the current assessment with previous ASAP and model. There is close agreement with the stock trends of the current assessment and the previous assessment. Mohn's Rho values were calculated for 5 retrospective runs 2018: 2013 for $\mathrm{F}_{\mathrm{bar}}(-0.034)$, SSB (0.155) and recruitment ( -0.266 ).

## State of the stock

Following a period of sustained decline, since 2008, SSB increased during 2010-2013. Since 2014, the SSB has increased markedly. The stock is characterized by highly variable recruitment. The model indicates above average recruitment for the 2009-2011 year class after below average recruitment for the 2007 and 2008 year classes. Recruitment in 2013 is amongst thehighestobserved in the time-series and has been followed by strongrecruitment in 2014 and 2015. The currentSSB is predicted to exceed any previously observed level.

### 8.4 Short-term projections

Short-term projections were performed using FLR libraries. Recruitment for 2019-2021 was estimated at (GM 1994-2016;357134 thousands). The F used in the forecast was derived as the F related to the TAC for 2019. The TAC for 2018 ( 3739 t ) was however, adjusted to account for the predicted landings that would be taken in rectangles 33 E 2 and 33 E 3 calculated as the average annual reallocation within the preceding ten years ( 430 t ) suggesting that landings in 2018 would be approximately 3309 equating to an $\mathrm{F}=0.253$.

Catches were split into landings and discards using the proportions of the catch that were discarded over the full the last three years. Input data for the short-term forecast are given in Table 11.7. The management options output is given in Table 11.9.

Estimates of the relative contribution of recent year classes to the 2020 landings and 2021 SSB are shown in Figure 11.12. As the very strong 2013 year class moves through the fishery the contribution to landings in 2020 is comprised of mainly the 2015 cohort (77\%), with the SSB in 2021 largely be dependent on a the 2016 cohort, comprising $62 \%$ of the biomass.

### 8.5 Biological reference points

## MSY evaluations

In response to an EU special request to provide plausible and updated FMSY ranges for Irish Sea haddock the management reference points for the stock were re-estimated (Table 11.10; ICES 2018). The Blim was set as the lowest SBB at which above recruitment in the upper quartile has been observed (2994t). The S-R plot for Irish Sea haddock shows no obvious S-R relationship
mainly because the recruitment is highly variable. Blim was estimated as 2994 t . MSY $\mathrm{B}_{\text {trigger }}$ is set to 4281 t as the stock has been fished at or below $\mathrm{F}_{\text {MSY }}$ for more than five years. FMSY median point estimates is 0.28 . The upper bound of the FMSY range giving at least $95 \%$ of the maximum yield was estimated to 0.35 and the lower bound at 0.20 . Flim is estimated to be 0.50 as F with $50 \%$ probability of $\mathrm{SSB}<\mathrm{B}_{\lim }$ with $\mathrm{F}_{\mathrm{pa}}$ as 0.38 calculated as $\mathrm{F}_{\text {lim }}$ combined with the assessment error; $F_{\lim } \times \exp (-1.645 \times \sigma) ; \sigma=0.20$.

## Yield and biomass-per-recruit

Not available for this stock, previous explorations are detailed in the stock annex.

### 8.6 Management plans

There is no specific management plan for haddock in the IrishSea. The regulations affecting Irish Sea haddock remain linked to those implemented under the cod management plan due to potential for bycatch of cod in a fishery targeting haddock (Council Regulation (EC) 1342/2008).

### 8.7 Uncertainties and bias in assessment and forecast

## Landings

Sampling levels of the landed catch for recent years are considered to be sufficient to support current assessment. However, within the assessment there is relocation of reported landings in rectangles 33 E 2 and 33 E 3 , which are not considered part of the stock. Historic misreporting estimates are considered in the assessment and accounted for, current misreporting is not considered to be a factor within the fishery.

## Discards

Sampling levels of discarding at sea remains high. For Northern Irish vessels targeting haddock $20.2 \%$ of trips are observed and $2.6 \%$ of the main Nephrops targeted fishery trips observed.

## Selectivity

A breakpoint in selectivity is applied in 2000, associated with management measures to reduce fishing mortality on cod. The model included three selectivity blocks in fishery-dependent data, reflecting bycatch and targeted fishery until the year 2000 (asymptotic). After 2007, a fleet selectivity pattern without targeted fishing of older fish (dome-shaped) is applied. During 2000-2007 a transition between a fully selected stock to a regime without targeted fishing of older fish is fitted. The use of current specified selectivity blocks may require review at annual at regular intervals. In the current assessment a new selectivity pattern for the fishery was added 20132018 with full selection of fish older than three years. With advice and management for haddock or other species, it is possible that the character of the fishery may change. A retrospective analysis demonstrated a consistent historic downward revision of the perceived SSB trend, however, there is consistent estimation of F . The initial two years of the retrospective plot show significant deviations. This was considered due to the model having a selectivity block, beginning in 2007, with reduced selection for older fish and the introduction of the UKFspW, with an asymptotic selectivity pattern, starting in 2007. The short period to estimate the selectivity parameters for both the fishery and survey index are considered to contribute to the instability of the model during this time.

## Surveys

The survey indices used in the model have spatial coverage of the assessment area. The combination of a recruitment index (NI-MIK), juvenile fish survey indices (NIGFS-WIBTS-Q1 \& NIGFS-WIBTS-Q4) and the UKFspW survey aimed at older fish using commercial fishing gear means that the full age range of the stock is covered by survey information.

### 8.8 Recommendations for next benchmark assessment

This stock was be benchmarked through the WKIrish process in 2016-2017.

### 8.9 References

EC. 2015. Commission Delegated Regulation (EU) 2015/2438 of 12 October 2015 establishing a discard plan for certain demersal fisheries in north-western waters.

Table 11.1. Landings ( $\mathbf{t}$ ) of HADDOCK in Division 7.a, 1984-present, as officially reported to ICES. (Working Group figures are given in Table 11.2).

| Country | 1984 |  | 1985 |  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 3 |  | 4 |  | 5 | 10 | 12 | 4 | 4 | 1 | 8 | 18 |
| France | 38 |  | 31 |  | 39 | 50 | 47 | n/a | n/a | n/a | 73 | 41 |
| Ireland | 199 |  | 341 |  | 275 | 797 | 363 | 215 | 80 | 254 | 251 | 252 |
| Netherlands | - |  | - |  | - | - | - | - | - | - | - | - |
| UK(E\&W) ${ }^{1}$ | 29 |  | 28 |  | 22 | 41 | 74 | 252 | 177 | 204 | 244 | 260 |
| UK (Isle of Man) | 2 |  | 5 |  | 4 | 3 | 3 | 3 | 5 | 14 | 13 | 19 |
| UK (N. Ireland) | 38 |  | 215 |  | 358 | 230 | 196 | ... | ... | ... | ... | ... |
| UK (Scotland) | 78 |  | 104 |  | 23 | 156 | 52 | 86 | 316 | 143 | 114 | 140 |
| Total | 387 |  | 728 |  | 726 | 1,287 | 747 | 560 | 582 | 616 | 703 | 730 |
| Country | 1994 |  | 1995 |  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| Belgium | 22 |  | 32 |  | 34 | 55 | 104 | 53 | 22 | 68 | 44 | 20 |
| France | 22 |  | 58 |  | 105 | 74 | 86 | n/a | 49 | 184 | 72 | 146 |
| Ireland | 246 |  | 320 |  | 798 | 1,005 | 1,699 | 759 | 1,238 | 652 | 401 | 229 |
| Netherlands | - |  | - |  | 1 | 14 | 10 | 5 | 2 | - | - | - |
| UK(E\&W) ${ }^{1}$ | 301 |  | 294 |  | 463 | 717 | 1,023 | 1,479 | 1,061 | 1,238 | 551 | 248 |
| UK (Isle of Man) | 24 |  | 27 |  | 38 | 9 | 13 | 7 | 19 | 1 | - | - |
| UK (N. Ireland) | ... |  | ... |  | ... | ... | ... | ... | ... | ... | ... | ... |
| UK (Scotland) | 66 |  | 110 |  | 14 | 51 | 80 | 67 | 56 | 86 | 47 | 31 |
| Total | 681 |  | 841 |  | 1,453 | 1,925 | 3,015 | 2,370 | 2,447 | 2,229 | 1,115 | 674 |
| Country |  | 2004 |  | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 |
| Belgium |  | 15 |  | 22 | 23 | 30 | 15 | 7 | 9 | 16 | 13 | 6.2 |
| France |  | 20 |  | 36 | 20 | 11 | 6 | 3 | 2 | 8 | 3 | . 7 |
| Ireland |  | 296 |  | 139 | 184 | 477 | 319 | 388 | 333 | 434 | 561 | 492 |
| Netherlands |  | - |  | - |  | - | - | - | - | - | - | - |
| UK (England \& Wales) ${ }^{1}$ |  | 421 |  | 344 | 419 | 559 | 521 | 446 | 593 | 355 | 236 | 154 |
| UK (Isle of Man) |  | - |  | - | - | - | 1 | 1 | - | - | <1 | <. 1 |
| UK (N. Ireland) |  | ... |  | ... | ... | ... | ... | ... | ... | ... | ... | ... |
| UK (Scotland) |  | 9 |  | 6 | 9 | 1 | 17 | 1 | 2 |  |  | - |
| United Kingdom |  |  |  |  |  |  |  |  |  |  | 236 | 154 |
| Total |  | 761 | 1 | 547 | 655 | 1078 | 879 | 846 | 939 | 813 | 813 | 654 |


| Country | 2014 | 2015 | 2016* | 2017 | 2018 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 7 | 7 | 5 | 5 | 4 |
| France | 0 | 7 | 1 | 5 | 0 |
| Ireland | 541 | 507 | 632 | 114 | 949 |
| Netherlands | - | - |  | - | - |
| UK (England \& Wales) ${ }^{1}$ | - | - | - | - |  |
| UK (Isle of Man) | $<1$ | $<1$ |  | - | - |
| UK (N. Ireland) | - | - |  | - | - |
| UK (Scotland) | 426 | 634 | 825 | 1240 | 1580 |
| United Kingdom | 974 | 1154 | 1463 | 2363 | 2532 |
| Total |  |  |  | - | - |

* Preliminary.
${ }^{1}$ 1989-2015 Northern Ireland included with England and Wales.
n/a = not available.

Table 11.2. Haddock in 7.a. Total international landings of haddock from the Irish Sea, 1972-present as officially reported to ICES. Working Group figures, assuming 1972-1992 official landings to be correct, are also given. The 1993-2005 WG estimates include sampled-based estimates of landings at a number of Irish Sea ports. Sample-based evidence confirms more accurate catch reporting since 2006. Landings in tonnes live weight. Since 1993, the landings have been corrected to exclude catches from the southernmost rectangles, which are not considered part of this stock.
$\left.\begin{array}{lccccc}\hline \text { Year } & \begin{array}{l}\text { Official land- } \\ \text { ings }\end{array} & \begin{array}{l}\text { WG land- } \\ \text { ings }\end{array} & \begin{array}{l}\text { ICES dis- } \\ \text { cards** }\end{array} & \begin{array}{l}\text { ICES } \\ \text { catch }\end{array} & \text { \% Discard }\end{array} \begin{array}{l}\text { Landings taken or reported in rec- } \\ \text { tangles 33E2 and 33E3 }\end{array}\right]$

| Year | Official landings | WG landings | ICES discards** | ICES <br> catch | \% Discard | Landings taken or reported in rectangles 33E2 and 33E3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 1925 | 3547 | 895 | 4442 | 20\% | 36 |
| 1998 | 3015 | 4874 | 1015 | 5889 | 17\% | 28 |
| 1999 | 2370 | 4095 | 634 | 4729 | 13\% | 34 |
| 2000 | 2447 | 1357 | 802 | 2159 | 37\% | 11 |
| 2001 | 2229 | 2246 | 269 | 2515 | 11\% | 74 |
| 2002 | 1115 | 1817 | 387 | 2204 | 18\% | 82 |
| 2003 | 674 | 659 | - | - | - | 64 |
| 2004 | 761 | 1217 | 392 | 1609 | 24\% | 53 |
| 2005 | 547 | 666 | 551 | 1217 | 45\% | 35 |
| 2006 | 655 | 633 | 306 | 939 | 33\% | 26 |
| 2007 | 1078 | 886 | 722 | 1608 | 45\% | 222 |
| 2008 | 879 | 786 | 643 | 1429 | 45\% | 194 |
| 2009 | 846 | 581 | 579 | 1160 | 50\% | 285 |
| 2010 | 939 | 679 | 508 | 1187 | 43\% | 267 |
| 2011 | 813 | 446 | 307 | 753 | 41\% | 374 |
| 2012 | n/a | 343 | 599 | 942 | 64\% | 473 |
| 2013 | 654 | 254 | 283 | 537 | 53\% | 410 |
| 2014 | 953 | 518 | 488 | 1006 | 49\% | 444 |
| 2015 | 1154 | 833 | 652 | 1451 | 44\% | 322 |
| 2016 | 1463 | 1008 | 298 | 1306 | 23\% | 455 |
| 2017 | 2363 | 1662 | 333 | 1995 | 17\% | 715 |
| 2018 | 2532 | 1993 | 568 | 2561 | 22\% | 532 |

Table 11.3. Haddock in 7.a: stock weights-at-age.

|  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 |
| 1993 | 0.02 | 0.095 | 0.42 | 1.043 | 1.759 | 2.563 |
| 1994 | 0.02 | 0.083 | 0.338 | 0.968 | 1.999 | 3.028 |
| 1995 | 0.02 | 0.085 | 0.347 | 0.785 | 1.708 | 3.219 |
| 1996 | 0.02 | 0.083 | 0.359 | 0.788 | 1.319 | 2.718 |
| 1997 | 0.022 | 0.07 | 0.357 | 0.863 | 1.435 | 2.391 |
| 1998 | 0.018 | 0.06 | 0.253 | 0.743 | 1.384 | 2.165 |
| 1999 | 0.016 | 0.057 | 0.226 | 0.561 | 1.294 | 2.262 |
| 2000 | 0.017 | 0.048 | 0.23 | 0.51 | 0.966 | 2.123 |
| 2001 | 0.018 | 0.051 | 0.201 | 0.548 | 0.93 | 1.822 |
| 2002 | 0.017 | 0.056 | 0.215 | 0.472 | 0.983 | 1.637 |
| 2003 | 0.017 | 0.05 | 0.229 | 0.485 | 0.798 | 1.52 |
| 2004 | 0.017 | 0.041 | 0.199 | 0.509 | 0.816 | 1.306 |
| 2005 | 0.018 | 0.031 | 0.165 | 0.459 | 0.902 | 1.347 |
| 2006 | 0.014 | 0.033 | 0.128 | 0.378 | 0.803 | 1.435 |
| 2007 | 0.019 | 0.034 | 0.136 | 0.299 | 0.68 | 1.402 |
| 2008 | 0.014 | 0.037 | 0.139 | 0.31 | 0.515 | 1.167 |
| 2009 | 0.025 | 0.042 | 0.153 | 0.326 | 0.563 | 0.98 |
| 2010 | 0.017 | 0.04 | 0.176 | 0.357 | 0.58 | 0.945 |
| 2011 | 0.018 | 0.052 | 0.167 | 0.407 | 0.624 | 0.937 |
| 2012 | 0.012 | 0.057 | 0.209 | 0.375 | 0.688 | 0.96 |
| 2013 | 0.023 | 0.059 | 0.233 | 0.491 | 0.673 | 1.115 |
| 2014 | 0.022 | 0.038 | 0.238 | 0.512 | 0.812 | 1.04 |
| 2015 | 0.017 | 0.046 | 0.153 | 0.577 | 0.97 | 1.371 |
| 2016 | 0.021 | 0.047 | 0.192 | 0.354 | 1.015 | 1.533 |
| 2017 | 0.022 | 0.054 | 0.137 | 0.347 | 0.809 | 1.476 |
| 2018 | 0.023 | 0.068 | 0.196 | 0.472 | 0.601 | 0.987 |

Table 11.4. Haddock in 7.a: Catch numbers-at-age.

|  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 |
| 1993 | 790 | 1568 | 2066 | 19 | 1 | 1 |
| 1994 | 16857 | 821 | 258 | 922 | 3 | 2 |
| 1995 | 950 | 8079 | 1587 | 107 | 220 | 5 |
| 1996 | 15171 | 1380 | 5510 | 728 | 16 | 30 |
| 1997 | 347 | 8828 | 1528 | 2388 | 201 | 16 |
| 1998 | 4209 | 4642 | 10532 | 252 | 488 | 42 |
| 1999 | 4944 | 3200 | 3436 | 4773 | 25 | 57 |
| 2000 | 287 | 11118 | 1771 | 466 | 457 | 418 |
| 2001 | 7883 | 425 | 3246 | 1074 | 30 | 89 |
| 2002 | 2105 | 8229 | 789 | 2063 | 142 | 18 |
| 2003 | 2000 | 2000 | 400 | 800 | 50 | 25 |
| 2004 | 10797 | 2056 | 421 | 827 | 46 | 78 |
| 2005 | 6048 | 4342 | 1416 | 285 | 193 | 34 |
| 2006 | 5334 | 2971 | 656 | 524 | 63 | 51 |
| 2007 | 2282 | 3537 | 3371 | 671 | 60 | 47 |
| 2008 | 2158 | 4569 | 2052 | 837 | 242 | 36 |
| 2009 | 4327 | 2490 | 2021 | 629 | 121 | 36 |
| 2010 | 3933 | 4058 | 834 | 464 | 309 | 59 |
| 2011 | 5669 | 2324 | 942 | 239 | 97 | 52 |
| 2012 | 6235 | 2799 | 774 | 201 | 27 | 28 |
| 2013 | 4525 | 1162 | 558 | 156 | 41 | 17 |
| 2014 | 1392 | 3854 | 1265 | 189 | 17 | 10 |
| 2015 | 518 | 1915 | 3087 | 324 | 63 | 5 |
| 2016 | 512 | 1845 | 907 | 1079 | 109 | 108 |
| 2017 | 231 | 783 | 2234 | 829 | 1096 | 78 |
| 2018 | 56 | 1039 | 5325 | 2845 | 426 | 526 |

Table 11.5. Haddock in 7.a: Available tuning data.

|  | H SEA hadd <br> S-WIBTS-Q1 <br> 2018 <br> 0.210 .25 | WG, AN | UNIN |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 10 | 139 | 569 | 31 | 0 | 0 |
| 1 | 10 | 644 | 58 | 183 | 0 | 0 |
| 1 | 10 | 24823 | 437 | 0.1 | 43 | 0 |
| 1 | 10 | 1065 | 3743 | 67 | 3 | 1.1 |
| 1 | 10 | 25118 | 474 | 1457 | 44 | 2.1 |
| 1 | 10 | 3913 | 8694 | 70 | 105 | 1.1 |
| 1 | 10 | 6058 | 680 | 2072 | 16 | 11.1 |
| 1 | 10 | 14028 | 1853 | 64 | 147 | 5 |
| 1 | 10 | 3277 | 6990 | 770 | 40 | 20.1 |
| 1 | 10 | 28755 | 842 | 1059 | 78 | 1.1 |
| 1 | 10 | 6966 | 14162 | 341 | 356 | 26.1 |
| 1 | 10 | 19945 | 2379 | 2206 | 45 | 35.1 |
| 1 | 10 | 24488 | 6454 | 406 | 234 | 15 |
| 1 | 10 | 13444 | 12721 | 2194 | 91 | 33.1 |
| 1 | 10 | 20918 | 11325 | 3661 | 240 | 27 |
| 1 | 10 | 7480 | 12009 | 2559 | 495 | 48.1 |
| 1 | 10 | 9345 | 3888 | 2877 | 163 | 42 |
| 1 | 10 | 17058 | 1765 | 524 | 239 | 27 |
| 1 | 10 | 17278 | 5543 | 299 | 67 | 50 |
| 1 | 10 | 13509 | 5266 | 1095 | 38 | 13 |
| 1 | 10 | 8245 | 5202 | 751 | 119 | 20 |
| 1 | 10 | 33807 | 2260 | 773 | 108 | 22 |
| 1 | 10 | 15495 | 22420 | 1297 | 407 | 44 |
| 1 | 10 | 14418 | 9109 | 5594 | 205 | 38 |
| 1 | 10 | 4321 | 18887 | 5524 | 323 | 33 |
| 1 | 10 | 7897 | 4683 | 7086 | 1709 | 1369 |


| $\begin{aligned} & \text { NIGFS-WIE } \\ & 19912018 \end{aligned}$ |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 110.83 | 0.88 |  |  |  |  |
| 04 |  |  |  |  |  |
| 1 | 36.127 | 0.716 | 3.965 | 0 | 0 |
| 1 | 2.042 | 151.766 | 1.171 | 0.959 | 0 |
| 1 | 15.289 | 101.536 | 0.753 | 0 | 0.045 |
| 1 | 1067.99 | 13.327 | 13.2 | 0.092 | 0.001 |
| 1 | 160.434 | 398.722 | 1.81 | 0.886 | 0.04 |
| 1 | 365.679 | 10.521 | 39.889 | 0.08 | 0.034 |
| 1 | 685.913 | 28.002 | 0.527 | 1.633 | 0.001 |
| 1 | 59.867 | 93.66 | 5.533 | 0.125 | 0.104 |
| 1 | 584.902 | 19.354 | 28.408 | 0.947 | 0 |
| 1 | 146.491 | 105.115 | 1.18 | 3.372 | 0 |
| 1 | 552.309 | 59.354 | 30.746 | 0.295 | 0.27 |
| 1 | 666.652 | 167.224 | 7.422 | 4.911 | 0.001 |
| 1 | 476.2 | 122.094 | 12.378 | 0.264 | 0.052 |
| 1 | 387.556 | 111.692 | 35.717 | 2.228 | 0.441 |
| 1 | 94.667 | 102.086 | 37.1 | 11.654 | 0.375 |
| 1 | 88.61 | 46.338 | 23.832 | 1.991 | 0.33 |
| 1 | 451.303 | 45.695 | 6.139 | 4.891 | 0.23 |
| 1 | 219.533 | 82.392 | 5.858 | 1.752 | 0.973 |
| 1 | 207.925 | 42.145 | 7.808 | 1.044 | 0.093 |
| 1 | 165.294 | 79.593 | 12.05 | 1.275 | 0 |
| 1 | 1004.22 | 8.279 | 1.531 | 0.179 | 0 |
| 1 | 339.218 | 311.607 | 68.768 | 3.016 | 0.423 |
| 1 | 455.385 | 81.189 | 108.663 | 2.309 | 0.362 |
| 1 | 99.046 | 154.865 | 52.207 | 4.273 | 0.281 |
| 1 | 191.946 | 42.885 | 90.324 | 15.934 | 6.202 |
| 1 | 690.663 | 167.338 | 12.891 | 16.507 | 2.003 |

NIMIK
$\begin{aligned} & 1994 \\ & 1\end{aligned} \quad 2018$
1
0 $0.38 \quad 0.47$

| 1 | 47000 |
| :---: | :---: |
| 1 | 1700 |
| 1 | 47800 |
| 1 | 14500 |
| 1 | 2500 |
| 1 | 15400 |
| 1 | 1700 |
| 1 | 17100 |
| 1 | 1200 |
| 1 | 4250 |
| 1 | 25970 |
| 1 | 8250 |
| 1 | 40240 |
| 1 | 3820 |
| 1 | 6638 |
| 1 | 18540 |
| 1 | 4532 |
| 1 | 6606 |
| 1 | 9818 |
| 1 | 28325 |
| 1 | 12892 |
| 1 | 48463 |
| 1 | 1800 |
| 1 | 26900 |
| 1 | 30954 |


| FSP Haddock: Tuning data |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 101 |  |  |  |  |  |  |
| UKFspW |  |  |  |  |  |  |
| 20052018 |  |  |  |  |  |  |
| 110.150 .25 |  |  |  |  |  |  |
| 05 |  |  |  |  |  |  |
| 1 | 0 | 0 | 1.774 | 1.506 | 4.981 | 0.291 |
| 1 | 0 | 0.308 | 7.749 | 7.336 | 0.546 | 1.115 |
| 1 | 0 | 0.208 | 42.727 | 37.286 | 6.289 | 0.697 |
| 1 | 0 | 0 | 4.657 | 12.836 | 7.213 | 0.794 |
| 1 | 0 | 0 | 0.662 | 3.99 | 1.443 | 0.541 |
| 1 | 0 | 0.627 | 1.422 | 3.78 | 2.753 | 0.866 |
| 1 | 0 | 0.048 | 0.598 | 1.976 | 1.121 | 0.81 |
| 1 | 0 | 0.27 | 4.135 | 4.772 | 0.79 | 0.226 |
| 1 | 0 | 0.035 | 3.684 | 7.674 | 1.742 | 0.176 |
| 1 | NA | NA | NA | NA | NA | NA |
| 1 | 0 | 0.437 | 31.2 | 19.349 | 5.051 | 0.554 |
| 1 | 0 | 0 | 0 | 59.769 | 12.592 | 6.205 |
| 1 | 0 | 0 | 19.748 | 85.536 | 246.488 | 10.838 |
| 1 | 0 | 0 | 0 | 36.397 | 62.861 | 55.448 |

Table 11.6. Haddock in 7.a: F-at-age.

|  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 |
| 1993 | 0.023 | 0.227 | 0.614 | 0.698 | 0.704 | 0.705 |
| 1994 | 0.024 | 0.236 | 0.639 | 0.726 | 0.733 | 0.734 |
| 1995 | 0.031 | 0.312 | 0.846 | 0.961 | 0.970 | 0.971 |
| 1996 | 0.025 | 0.250 | 0.675 | 0.767 | 0.775 | 0.775 |
| 1997 | 0.031 | 0.312 | 0.844 | 0.959 | 0.968 | 0.969 |
| 1998 | 0.033 | 0.327 | 0.886 | 1.007 | 1.016 | 1.017 |
| 1999 | 0.048 | 0.483 | 1.308 | 1.486 | 1.500 | 1.501 |
| 2000 | 0.034 | 0.342 | 0.926 | 1.053 | 1.063 | 1.063 |
| 2001 | 0.124 | 0.422 | 0.753 | 0.803 | 0.562 | 0.401 |
| 2002 | 0.159 | 0.539 | 0.961 | 1.026 | 0.717 | 0.513 |
| 2003 | 0.128 | 0.435 | 0.776 | 0.828 | 0.579 | 0.414 |
| 2004 | 0.123 | 0.416 | 0.743 | 0.792 | 0.554 | 0.396 |
| 2005 | 0.100 | 0.341 | 0.609 | 0.649 | 0.454 | 0.325 |
| 2006 | 0.057 | 0.195 | 0.347 | 0.371 | 0.259 | 0.185 |
| 2007 | 0.095 | 0.324 | 0.578 | 0.616 | 0.431 | 0.308 |
| 2008 | 0.144 | 0.512 | 0.578 | 0.545 | 0.299 | 0.145 |
| 2009 | 0.111 | 0.393 | 0.444 | 0.419 | 0.230 | 0.112 |
| 2010 | 0.162 | 0.574 | 0.648 | 0.611 | 0.335 | 0.163 |
| 2011 | 0.085 | 0.302 | 0.341 | 0.322 | 0.176 | 0.086 |
| 2012 | 0.088 | 0.314 | 0.354 | 0.334 | 0.183 | 0.089 |
| 2013 | 0.009 | 0.064 | 0.112 | 0.107 | 0.112 | 0.112 |
| 2014 | 0.011 | 0.086 | 0.150 | 0.144 | 0.150 | 0.150 |
| 2015 | 0.010 | 0.077 | 0.134 | 0.128 | 0.134 | 0.134 |
| 2016 | 0.007 | 0.053 | 0.093 | 0.089 | 0.093 | 0.093 |
| 2017 | 0.009 | 0.068 | 0.118 | 0.113 | 0.118 | 0.118 |
| 2018 | 0.012 | 0.089 | 0.156 | 0.149 | 0.156 | 0.156 |

Table 11.7. Forecast input data.

| Variable | Value | Source | Notes |
| :--- | :--- | :--- | :--- |
| F ages 2-4 (2019) | 0.253 | ICES <br> (2019a) | F in 2017 predicted for TAC, adjusted for annual reallocation of land- <br> ings from rectangles 33E2 and 33E3 (ten year average value) |
| SSB (2020) | 14160 | ICES <br> (2019a) | Short-term forecast |
| Rage 0 (2019 and 2020) <br> (thousand) | 357134 | ICES <br> (2019a) | Geometric mean (2006-2016) |
| Catch (2019) | 3746 | ICES <br> (2019a) | Short-term forecast |
| Wanted catch * (2019) | 3310 | ICES <br> (2019a) | Average discard pattern (2016-2018) |
| Unwanted catch <br> (2019) | 437 | ICES <br> (2019a) | Average discard pattern (2016-2018) |

* "Wanted catch" is used to describe fish that would be landed in the absence of the EU landing obligation.

Table 11.8. Haddock in Division7.a. Assessment results. All weights are in tonnes. Low and high refer to 1 std confidence limits.

| Year | Recruitment | High | Low | SSB | High | Low | Landings <br> tonnes | Discards <br> tonnes | F High <br> Ages 2-4 |  | Low |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 0 |  |  |  |  |  |  |  |  |  |  |
|  | thousands |  |  | tonnes |  |  |  |  |  |  |  |
| 1993 | 153743 | 172355 | 135131 | 2391 | 2725 | 2058 | 813 | 365 | 0.67 | 0.81 | 0.53 |
| 1994 | 527015 | 578308 | 475722 | 2234 | 2631 | 1838 | 1042 | 468 | 0.70 | 0.86 | 0.54 |
| 1995 | 62874 | 75047 | 50701 | 2376 | 2810 | 1941 | 1736 | 780 | 0.93 | 1.14 | 0.71 |
| 1996 | 1357285 | 1488285 | 1226285 | 4879 | 5497 | 4262 | 2981 | 709 | 0.74 | 0.88 | 0.60 |
| 1997 | 210990 | 241958 | 180022 | 4066 | 4758 | 3374 | 3547 | 895 | 0.92 | 1.11 | 0.74 |
| 1998 | 344353 | 386475 | 302231 | 8209 | 9089 | 7329 | 4874 | 1015 | 0.97 | 1.12 | 0.82 |
| 1999 | 677646 | 745127 | 610165 | 5648 | 6411 | 4885 | 4095 | 634 | 1.43 | 1.68 | 1.18 |
| 2000 | 98299 | 115777 | 80821 | 2727 | 3195 | 2260 | 1357 | 802 | 1.01 | 1.24 | 0.79 |
| 2001 | 709368 | 784132 | 634604 | 3873 | 4440 | 3305 | 2246 | 269 | 0.71 | 0.83 | 0.58 |
| 2002 | 134462 | 155680 | 113244 | 2938 | 3454 | 2421 | 1817 | 387 | 0.90 | 1.08 | 0.72 |
| 2003 | 424040 | 480565 | 367515 | 3518 | 4085 | 2951 | 1517 | 390 | 0.73 | 0.88 | 0.57 |
| 2004 | 646527 | 719493 | 573561 | 2636 | 3184 | 2088 | 1217 | 392 | 0.70 | 0.86 | 0.54 |
| 2005 | 496007 | 550841 | 441173 | 2547 | 3052 | 2041 | 666 | 551 | 0.57 | 0.71 | 0.44 |
| 2006 | 566750 | 622671 | 510829 | 3288 | 3853 | 2723 | 633 | 306 | 0.33 | 0.40 | 0.25 |
| 2007 | 224231 | 250795 | 197667 | 4298 | 4926 | 3669 | 886 | 722 | 0.54 | 0.65 | 0.44 |
| 2008 | 158495 | 179166 | 137824 | 4416 | 5103 | 3730 | 786 | 643 | 0.47 | 0.57 | 0.38 |
| 2009 | 336811 | 374765 | 298857 | 3848 | 4571 | 3125 | 581 | 579 | 0.36 | 0.44 | 0.29 |
| 2010 | 252631 | 283537 | 221725 | 3344 | 4027 | 2661 | 679 | 508 | 0.53 | 0.65 | 0.42 |
| 2011 | 314140 | 352344 | 275936 | 3037 | 3697 | 2376 | 446 | 307 | 0.28 | 0.34 | 0.22 |
| 2012 | 308690 | 352445 | 264935 | 3345 | 4028 | 2662 | 343 | 599 | 0.29 | 0.35 | 0.23 |
| 2013 | 1458723 | 1627793 | 1289653 | 4287 | 5135 | 3439 | 254 | 282 | 0.110 | 0.135 | 0.085 |
| 2014 | 648601 | 736419 | 560783 | 5958 | 6982 | 4934 | 518 | 488 | 0.148 | 0.179 | 0.116 |
| 2015 | 943094 | 1075454 | 810734 | 12058 | 13782 | 10335 | 833 | 652 | 0.132 | 0.159 | 0.105 |
| 2016 | 287929 | 337962 | 237896 | 15894 | 18152 | 13636 | 1008 | 298 | 0.091 | 0.110 | 0.073 |
| 2017 | 279938 | 335044 | 224832 | 20049 | 22931 | 17168 | 1662 | 333 | 0.117 | 0.141 | 0.092 |
| 2018 | 281126 | 367573 | 194679 | 20241 | 23324 | 17159 | 1993 | 568 | 0.154 | 0.188 | 0.120 |
| 2019 | 357134* |  |  | 18945 |  |  |  |  |  |  |  |

Table 11.9 Haddock in Division 7.a. Annual catch scenarios. All weights are in tonnes.

| Basis | Total catch (2020) | Wanted catch * (2020) | Unwanted catch * (2020) | $\begin{aligned} & F_{\text {total }} \\ & \text { (2020) } \end{aligned}$ | $\begin{aligned} & F_{\text {wanted }} \\ & (2020) \end{aligned}$ | $\begin{aligned} & \text { Funwanted } \\ & (2020) \end{aligned}$ | $\begin{aligned} & \text { SSB } \\ & \text { (2021) } \end{aligned}$ | $\begin{aligned} & \text { \% SSB } \\ & \text { change } \\ & \text { ** } \end{aligned}$ | \% Advice change^ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |  |
| EU MAP ^^: FMSY | 3156 | 2667 | 489 | 0.28 | 0.19 | 0.086 | 10493 | -26 | -15.6 |
| F = MAP F MSY $_{\text {lower }}$ | 2333 | 1975 | 358 | 0.20 | 0.14 | 0.061 | 11317 | -20 | -38 |
| F = MAP $\mathrm{F}_{\text {MSY }}^{\text {upper }}$ | 3830 | 3232 | 598 | 0.35 | 0.24 | 0.11 | 9824 | -31 | 2.4 |
| $\mathrm{F}=0$ | 0 | 0 | 0 | 0 | 0 | 0 | 13678 | -3.4 | -100 |
| $\mathrm{F}_{\mathrm{pa}}$ | 4106 | 3463 | 644 | 0.38 | 0.26 | 0.12 | 9550 | -33 | 9.8 |
| Flim | 5141 | 4324 | 817 | 0.50 | 0.35 | 0.15 | 8533 | -40 | 38 |
| $\mathrm{SSB}_{2020}=\mathrm{Blim}$ | 11133 | 9142 | 1991 | 1.64 | 1.14 | 0.50 | 2994 | -79 | 198 |
| $\mathrm{SSB}_{2021}=\mathrm{B}_{\mathrm{pa}}$ | 9788 | 8099 | 1690 | 1.28 | 0.89 | 0.39 | 4160 | -71 | 162 |
| $\mathrm{SSB}_{2021}=$ MSY Briger | 9653 | 7992 | 1661 | 1.24 | 0.86 | 0.38 | 4281 | -70 | 158 |
| $\mathrm{F}=\mathrm{F}_{2019}$ | 2887 | 2441 | 446 | 0.25 | 0.18 | 0.07 | 10762 | -24 | -23 |

* "Wanted" and "unwanted" catch are used to describe fish that would be landed and discarded in the absence of the EU landing obligation, based on discard rate estimates for 2016-2018.
** SSB 2021 relative to SSB 2020.
*** Total catch in 2020 relative to TAC in 2019 ( $\mathbf{3 7 3 9}$ tonnes).
^Advice value for 2020 relative to advice value for 2019 ( $\mathbf{3 7 3 9}$ tonnes).
Please check years are correct.
$\wedge \wedge$ EU multiannual plan (MAP) for the Western Waters (EU, 2019).

Table 11.10. Haddock in 7.a Management reference points.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 4281 tonnes | 5th percentile of BMSY; Irish Sea haddock has been fished at, or below $\mathrm{F}_{\text {MSY }}$ for >five years. | ICES <br> (2018a) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.28 | Median point estimates of EqSim with segmented regression stock-recruitment relationship | ICES <br> (2018a) |
|  | $\mathrm{F}_{\text {MSYLower }}$ | 0.20 | F at 95\% of MSY below $\mathrm{F}_{\text {MSY }}$ | ICES <br> (2018a) |
|  | $\mathrm{F}_{\text {MSYupper }}$ | 0.35 | F at 95\% of MSY above $\mathrm{F}_{\text {MSY }}$ | ICES <br> (2018a) |
| Precautionary approach | $\mathrm{Bl}_{\text {lim }}$ | 2994 tonnes | Lowest observed SSB with $>$ 75th percentile recruitment | ICES <br> (2018a) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 4160 tonnes | $\mathrm{B}_{\text {lim }}$ combined with the assessment error; $\mathrm{B}_{\text {lim }} \times$ $\exp (1.645 \times \sigma) ; \sigma=0.20$ | ICES <br> (2018a) |
|  | $\mathrm{F}_{\text {lim }}$ | 0.50 | F with $50 \%$ probability of SSB $<\mathrm{B}_{\text {lim }}$ | ICES <br> (2018a) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.38 | $\mathrm{F}_{\text {lim }}$ combined with the assessment error; $\mathrm{F}_{\text {lim }} \times$ $\exp (-1.645 \times \sigma) ; \sigma=0.2$ | ICES <br> (2018a) |
| Management plan | SSB ${ }_{\text {MGT }}$ | Not applicable |  |  |
|  | $\mathrm{F}_{\text {MGT }}$ | Not applicable |  |  |



Figure 11.1. Haddock in 7.a: Growth of haddock in the Irish Sea. Top two panels: mean length-at-age in UK(NI) groundfish surveys in March (NIGFS-WIBTS-Q1), by y ear and age, and expected mean weight-at-length based on length-weight parameters from each survey. Lower panels: mean length-at-age from March surveys, and from Quarter 1 commercial landings at-age 3 and over, by year class. Lines are von Bertalanffy model fits with year-class effect included. Model residuals are shown for the fit without year-class effects, and for the fit with y ear-class effects.


Figure 11.2. Haddock in 7.a: Trends in log-standardised survey indices.

## UKFspW

|  |  | age 4 vs 5 |
| :---: | :---: | :---: |
|  |  |  |
| $\begin{aligned} & \times \\ & \stackrel{\times}{0} \\ & . \overline{=} \\ & \text { ס } \end{aligned}$ | age 3 vs 4 | age 3 vs 5 |
|  |  |  |
| age 2 vs 3 | age 2 vs 4 | age 2 vs 5 |
|  |  |  |

NIGFS-WIBTS-Q4


## NIGFS-WIBTS-Q1



Figure 11.3. Haddock in 7.a: Scatterplot matrix of $\log$ indices of cohorts at different ages.


Figure 11.4 Residuals from fitted and observed catch age proportions from update assessment as applied in 2018 (A) and with new starting estimate for selectivity of 0 age haddock from 2013-2018 (B).


Figure 11.5. Fitted and observed catch from update assessment.


Figure 11.6. Fitted and observed catch age proportions from update assessment.


Figure 11.7. Observed catch numbers 2005-present.


Figure 11.8. Fitted and observed index series from update assessment.


Figure 11.9. Fitted and observed index age proportions from update assessment.


Figure 11.10. A retrospective plot the final update model.




Figure 11.11. Haddock in Division7.a. Historical assessment results.

Landings yield 2020


Figure 11.12. Haddock in 7a. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these y ear classes.

## 9 Haddock in divisions 7.b,c,e-k

## Type of assessment in 2019

The 2019 assessment followed the Stock Annex 5 procedure performed in the preceding year.

## ICES advice applicable to 2018

Last year's full advice is available in the ICES Advice 2018, Book 5. The headline advice was as follows:
"For haddock in Division 7.b-k, ICES advises that when the EU multiannual plan (MAP) is applied, catches in 2020 that correspond to the F ranges in the MAP are between 11418 and 23262 tonnes. According to the MAP, catcheshigher than those corresponding to $F_{M S Y}(16671$ tonnes) can only be taken under conditions specified in the MAP, whilst the entire range is considered precautionary when applying the ICES advice rule."

### 9.1 General

## Stock description and management units



The basis for the stock assessment area $7 . b, c, e-k$ is described in detail in the stock annex.
Figure 7.4.1 shows the spatial distribution of international haddock landings in the NE Atlantic for 2016. It is clear from the figure that the stock extends into Area 8 and it could be argued that landings from 8 should be included in the stock area. In recent years these landings varied between 20 and 300 t which is up to $4 \%$ of the total landings in the stock area.
The TAC for haddock is set for the combined Areas $7 . b-k, 8,10$ and 10 and EU waters of CECAF 34.1.1. This does not correspond to the stock assessment area (7.b-k).

2018 management (Council Regulation (EU) 2018/120)

| Species: | Haddock <br> Melanogrammus acglefinus | Zone:7b-k, 8,9 and 10; Union waters of CECAF 34.1.1 <br> (HAD/7X7A34) |  |
| :--- | ---: | :--- | :--- |
| Belgium | 77 |  |  |
| France | 4606 |  |  |
| Ireland | 1536 |  |  |
| United Kingdom | 691 | 6910 | Analytical TAC <br> Article 7(2) of this Regulation applies <br> Article 12(1) of this Regulation applies |
| TAC | 6910 |  |  |

2019 management (Council Regulation (EU) 2019/124)

| Species: | Haddock <br> Melanogrammus acglefinus | Zone:7b-k, 8,9 and 10; Union waters of CECAF 34.1.1 <br> (HAD/7X7A34) |  |
| :--- | ---: | :--- | :--- |
| Belgium | 93 |  |  |
| France | 5552 |  |  |
| Ireland | 1851 |  |  |
| United Kingdom | 833 | 8329 | Analytical TAC <br> Article 7(2) of this Regulation applies <br> Union |
| TAC | 8329 |  |  |
|  |  |  |  |

Since 2009, a separate TAC is set for 7.a haddock; previously a separate allocation for 7.a existed within the TAC for $7,8,9$ and 10 .

During the 2011 December fisheries councilmeeting, Ireland, UK and France agreed to introduce additional technical measures to reduce the high levels of gadoids discards recently observed in the Celtic Seas. In consultation with national governments and the NWWRAC it was agreed to introduce the mandatory use of a 110 mm square mesh panel in Nephrops trawls and a 100 mm panel in gadoid fisheries. While the regulation was not introduced until 14th August 2012 (EC Regulation 737/2012), it is understood that for both French and Irish fleets, the technical measures were in practice introduced much earlier in the year by the national administrations.

### 9.2 The fishery

The official landings reported to ICES and Working Group estimates of the landings and discards are given in Table 7.4.1. The historic landings are also shown in Figure 7.4.2.

Before 2002, the TAC was well in excess of the landings in the TAC area (Table 7.4.1a). The TAC appeared to become restrictive for France in 2003-2004 and Ireland in 2001-2003 and perhaps after (Table 7.4.1a and Figure 7.4.2b). (WGSSDS05 provided some qualitative evidence that misreporting was a problem). During 2005-2008 landings were well below the TAC. In 2009 and 2010 the total landings were still below the TAC, but the quota appeared to become restrictive again for Ireland and Belgium. Since 2011 the TAC has been close to the total landings and can be assumed to be restrictive for all countries.

Figure 7.4.2a gives a long-term overview of the landings of haddock. The time-series is characterized by a number of peaks with rapid increases in the landings, mostly followed by rapid
decreases within a few years, suggesting the fishery was taking advantage of sporadic events of very high recruitment. During the 1960s and 1970s, three such peaks in landings occurred: the landings increased from less than 4000 t to 10000 t or more. During the 1980 s and early 1990s, landings were relatively stable around 2000-4000 t . During the mid-1990s the haddock landings increased again to over 10000 t , mirroring increased landings in the Irish Sea in that period. Since the late 1990s the landings have varied between 7000 and $10000 t$ and in 2012, the landings were the highest on record at more than 18000 t .
The discard estimate for 2010 was the highest on record at 16547 tonnes (Table 7.4.1b), this was mainly a consequence of the 2009 cohort entering the fishery.
Table 7.4.2 and Figure 7.4.3 show that Irish commercial lpue was relatively low between 2003 and 2007 after which it increased. Effort in the French gadoid fleet has declined considerably since theearly 2000s as the result of a decommissioning scheme. The French and Irish7.fgh fleets both showed an increase in lpue as the strong 2009 cohort entered the fishery. These data are presented for auxiliary information only; these fleets are not used directly in the assessment.

### 9.2.1 Information from the Industry

The French and Irish fishing industry have reported that the abundance and distribution of haddock increased in 2016 and that this continued into 2018. Due to the restrictive TAC the industry have reported to national scientists that there is increased discarding of haddock.

### 9.3 Data

### 9.3.1 Landings and discard numbers-at-age

Discard and retained catch-age distributions for 2013 and 2018 are shown in Figure 7.4.4. Many of the discarded fish will be above the MLS, which is likely to be the result of restrictive quota.

Figures 7.4.5 and 7.4.6 shows the available time-series of catch (discards and retained catch) and age distributions.

The historic approach to raising the catch numbers-at-age is given in the stock annex. France and Ireland had allocated age distributions to most unsampled catches before uploading to InterCatch. For métiers where discards were not provided, the discards were estimated from the discard rate of métiers that had both landings and discards. Sampled and unsampled métiers, by country and age class, are shown in Figure 7.4.7.

Landings numbers-at-age are given in Table 7.4.3a and discard numbers-at-age are given in Table 7.4.3b. Despite some uncertainty about the quality of the discard data, it is possible to track strong year classes in both the discards and the landings-at-age matrices. Discards account for a large proportion of the catch numbers up to age 3. Figure 7.4 .6 shows the proportions-at-age that are discarded; over the last ten years $97 \%$ of one year-olds, $80 \%$ of two year-olds and $40 \%$ of three-year olds have been discarded. By number, $77 \%$ of the total catch was discarded ( $46 \%$ by weight; average last ten years). There is a trend for increasing proportions of two and 3-year olds to be discarded, in the mid-nineties around half of the 2-year olds were discarded and around $10 \%$ of 3-year olds while in recent years around $80 \%$ of 2 -year olds and $30 \%$ of 3-year olds were being discarded.
Catch and stock weights-at-age are given in the ASAP input file (Table 7.4.4). Figure 7.4 .8 shows that the raw stock weights-at-age which are fairly noisy, a 3-year running average was applied to the stock weights used in the assessment. There appear to be cyclical trends in the weights-atage that follow cohorts (rather than year-effects).

### 9.3.2 Biological

The assumptions of natural mortality and maturity are described in the stock annex. The maturity ogive used in the assessment is knife-edged at-age 2. Recent Irish maturity data from 20042014 (working document to WGCSE15) suggested a similar maturity ogive for females, but also indicated that a significant number of males mature before the age of two.

### 9.3.3 Surveys and commercial tuning fleets

The available surveys and commercial tuning fleets are described in detail in the stock annex. One survey index is used in the assessment: the FR-IRL-IBTS index, which is a combined index from the French EVHOE Q4 WIBTS and Irish IGFS Q4 WIBTS surveys. Additionally one commercial tuning fleet is used: the IR-GAD index, which is the Irish gadoid fleet in selected rectangles of 7.gj. The index data are given in the ASAP input file(Table7.4.4). The standardised indices are given by year in Figure 7.4.9 and by cohort in Figure 7.4.10. Figure 7.4.11 shows the scatterplot matrices of the log indices. These plots indicate that the internal consistency of the indices is robust. The IR-GAD index (Figure 7.4.9) shows an increasing trend over time, mainly as a result of the relatively strong 2002 and 2009 cohorts.

### 9.4 Historical stock development

Model used: ASAP; (XSA is also used for quality control purposes).
Software used: ASAP V3.0.17 NOAA Fisheries toolbox (http://nft.nefsc.noaa.gov)
FLR with R version 3.1.2 with packages FLCore 2.5.20150116, FLAssess _2.5.20130716, FLXSA 2.5. 20140808 and FLEDA 2.5 (http://flr-project.org/)

### 9.4.1 Data screening

The general approach to data screening and analysis was followed in addition to the data exploration tools available in the FLR package FLEDA. The results of the data screening are fully documented using R markdown and are available in the folder 'Data $\backslash$ Stock $\backslash$ had- $7 \mathrm{bce}-\mathrm{k}$ ' on the ICES SharePoint.

### 9.4.2 Final update assessment

The final assessment was run with the same settings as established by WKROUND 2012 and described in the stock annex. Discards were combined with the landings and not supplied separately to the model.

Figure 7.4 .12 shows the residuals of that catch proportions-at-age. For age classes where discards dominate, the residuals are relatively large. There is no obvious pattern in the younger ages but the residuals in the older ages at the start of the time-series are mostly positive. The 2010 year class has the strongest residuals in later year (ages of 4 and above) with the signal continuing with the year class as they age, indicating that the model does not 'believe' the 2010 cohort is as strong as the index suggests. Theobserved and predicted catches are shown in Figure 7.4.12. The predicted catches were generally slightly lower than observed in most recent years while they were generally higher than observed from 2002-2006.

In the proportions-at-age residual plots of the survey (FRA-IRL-IBTS) commercial (IRL-GAD) indices (Figure 7.4.14) there are no consistent patterns, with only minor differences between observed and predicted values. The observed and predicted index cpue values are shown in Figure 7.4.15. The model closely follows the survey index, though in 2012 and 2016 the predicted index in the IRL-GAD index was above the observed, as the assessment balanced the observed catch in the FRA-IRL scientific survey indices. The scientific survey index also observed a notable increase in the 2018 survey, which follows through the assessment as a notable 2018 recruitment event, and is shown in Figure 7.4.14 to comprise age 0 recruits.
The selectivity of the catch data was freely estimated for ages 1 and 2 by the model. For the other ages, selectivity was fixed. Table 7.4 .5 shows the model estimates for ages 1 and 2 . Selectivity of the FR-IR-IBTS index was fixed at 1 for all ages that were included and selectivity (exploratory data analysis shows that log catch numbers of those ages decline in straight lines) of the IRLGAD index was freely estimated for age 3 and fixed at one for older ages. (Discards are not included in this commercial fleet therefore selectivity was not assumed to be the same as that of the catch data).

The ASAP assessment is shown in Figure 7.4.16, detailing catch, landings, SSB F and recruits with stdev.

### 9.4.3 Comparison with previous assessments

Figure 7.4.16 shows the comparison of the current assessment with an XSA assessments. The two assessment models mirror catch, SSB and recruitment well, with estimates of F being of a similar range if dissimilar in precision in the recent time period.

### 9.4.4 State of the stock

Table 7.4.6 shows the estimated fishing mortality-at-age and Table 7.4 .7 shows the stock num-bers-at-age. The stock summary is given in Table 7.4.8.

The spawning-stock biomass (SSB) peaked in 2011 as the very strong 2009 year class matured; this cohort was followed by three years of below-average recruitment which led to a rapid decline in SSB after 2011. Recent recruitment has varied around the average, with a notable peak in 2009 and in 2018. SSB appears to have stabilised, while fishing mortality ( F ) has been above $\mathrm{F}_{\text {MSY }}$ for the entire time-series but shows a declining trend.

### 9.5 Short-term projections

Because recruitment of haddock is characterised by sporadic events, the assumed geometric mean (GM) recruitment for the intermediate year introduces significant uncertainty for the SSB estimate in 2019. The short-term predictions are expected to give a reasonably reliable estimate of landings in 2019 (assuming average F 2015-2017), which are largely based on the estimates of the 2016 and 2018 recruitments. In the past, recruitment has generally been accurately estimated.

Short-term projections were performed using FLR libraries. Recruitment for 2019-2021 was estimated at 252713 (GM 1993-2016; thousands). Three year averages were used for F (unscaled) and weights-at-age. Catches were split into landings and discards using the proportions of the catch that were discarded over the full time-series. This was done because the discard pattern over the last four years are unlikely to persist: the proportion of discards in the 2013-2014 was considerably lower than the historic proportion of discards.

Input data for the short-term forecast are given in Table 7.4.9. The single option output is given in Tables 7.4.10 and 7.4.11 gives the management options.

Intermediate year assumptions for the catch advice were as follows:

| Variable | Value | Notes |
| :--- | :--- | :--- |
| $\mathrm{F}_{\text {ages 3-5 (2019) }}$ | 0.77 | $\mathrm{~F}_{\mathrm{sq}}=\mathrm{F}_{\text {Average }}$ 2015-2018), rescaled to 2018 |
| SSB (2020) | 49821 | tonnes; $\mathrm{F}_{\mathrm{sq}}=0.77$ |
| $\mathrm{R}_{\text {age }} 0$ (2019-2020) | 252713 | thousands; Geometric mean (1993-2016) |
| Catch (2019) | 20457 | tonnes; $\mathrm{F}_{\mathrm{sq}}=0.77$ |
| Wanted catch (2019) | 6963 | tonnes; Average discard rate (1993-2018) |
| Unwanted catch (2019) | 13494 | tonnes; Average discard rate (1993-2018) |

Estimates of the relative contribution of recent year classes to the 2020 landings and 2021 SSB are shown in Figure 7.4.17. The extremely high recruitment observed in 2018 accounts for over 77\% of the projected landings in 2020 while the GM 2019 assumption only accounts for $3 \%$ of 2020 landings. The 2018 cohort also contributes largely to the estimated SSB in 2021 ( $58 \%$ ) with a second, large proportion coming from the 2019 GM assumption. At GM recruitment and status quo $F$, SSB will remain well above $B_{\text {trigger }}$.

### 9.6 MSY evaluations and biological reference points

ICES carried out and evaluation of MSY and PA reference points for this stock at WKMSYREF4 (ICES, 2016a). The results have been published earlier this year (ICES, 2016b) are summarized below:

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 10000 t | $\mathrm{B}_{\mathrm{pa}}$ | ICES <br> (2016a) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.40 | Median point estimates of EqSim with a segmented regression stock-recruitment relationship. | ICES <br> (2016a) |
| Precautionary approach | $\mathrm{Bl}_{\text {lim }}$ | 6700 t | Lowest observed SSB | ICES <br> (2016a) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 10000 t | $\mathrm{B}_{\text {lim }}$ combined with the assessment error; $\mathrm{B}_{\text {lim }} \times \exp (1.645 \times \sigma)$; $\sigma=0.26$ | ICES <br> (2016a) |
|  | $\mathrm{F}_{\text {lim }}$ | 1.41 | F with $50 \%$ probability of $\mathrm{SSB}<\mathrm{B}_{\text {lim }}$ | ICES <br> (2016a) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.89 | $F_{\text {lim }}$ combined with the assessment error; $\mathrm{F}_{\mathrm{lim}} \times \exp (-1.645 \times \sigma)$; $\sigma=0.28$ | ICES <br> (2016a) |
| Management plan* | MAP <br> MSY $B_{\text {trigger }}$ | 10000 t | MSY $\mathrm{B}_{\text {trigger }}$ |  |
|  | MAP $\mathrm{Bl}_{\text {lim }}$ | 6700 t | $\mathrm{Blim}_{\text {lim }}$ |  |
|  | MAP $\mathrm{F}_{\text {MSY }}$ | 0.40 | $\mathrm{F}_{\text {MSY }}$ |  |
|  | MAP range <br> $\mathrm{F}_{\text {lower }}$ | 0.26 | Consistent with ranges provided by ICES (2016b), resulting in no more than $5 \%$ reduction in long-term yield compared with MSY. |  |
|  | MAP range <br> $\mathrm{F}_{\text {upper }}$ | 0.60 | Consistent with ranges provided by ICES (2016b), resulting in no more than $5 \%$ reduction in long-term yield compared with MSY. |  |

### 9.7 Management plans

The EU multiannual plan (MAP) for the Western Waters (EU, 2019), incorporating the stockhaddock $7 . \mathrm{b}, \mathrm{c}, \mathrm{e}-\mathrm{k}$ has been agreed. This MAP "establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks", under article 17 states that "It is appropriate to establish the target fishing mortality ( F ) that corresponds to the objective of reaching and maintaining MSY as ranges of values which are consistent with achieving MSY(FMSY). Those ranges, based on best available scientific advice, are necessary in order to provide flexibility to take account of developments in the scientific advice, to contribute to the implementation of the landing obligation and to take into account the characteristics of mixed fisheries."

### 9.8 Uncertainties and bias in assessment and forecast

### 9.8.1 Landings

Sampling levels of the landed catch for recent years are considered to be sufficient to support current assessment approaches, although the assessment is contingent on the accuracy of the landings statistics.

### 9.8.2 Discards

Irish discards have been monitored since 1995. The number of trips sampled has varied considerably over time (between three and 62 trips per year). Sample numbers were particularly low in 1995, 1999-2002 and 2006. During the remaining years, the number of sampled trips was considered sufficient to give reliable estimates of discards.

French discard data exist from 2004 onwards but the data are not considered to be reliable before 2008. The time-series of French discards was reconstructed by assuming that $90 \%$ of one-year olds, $50 \%$ of two-year olds and $10 \%$ of three year olds were discarded throughout the time-series. These proportions were estimated from the available discard and retained catch data provided by France. Because French discards are estimated to account for $80-86 \%$ of the international discards (by weight; 2008-2012), there is considerable uncertainty around the historic discard estimates. However WKROUND (2012) concluded that the ASAP assessment is relatively robust to the uncertainty in the discard estimates.

Although recent discard estimates are considered to be more reliable, the problem remains that the number of observer trips is very small compared to the total number of trips (typically $<1 \%$ of all trips are sampled). The level of uncertainty owing to the small sample sizes is likely to be high but the cost of increasing discard coverage would be considerable.

### 9.8.3 Selectivity

As a consequence of the introduction of square-mesh panels in the Celtic Sea, the selectivity of the fleet might be expected to change. The regulations were introduced in the second half of 2012 (although many vessels had already voluntarily fitted panels earlier that year). STECF (PLEN-13-03) investigated the efficiency of the introduction of the square-mesh panel in the Celtic Sea and did not find evidence for a change in selectivity in 2012 or 2013. A possible change in selectivity was investigated using a number of different approaches:

- There is no evidence of a 'block' of negative residuals of young fish in recent years from the catch proportions-at-age residuals (Figure 7.4.12).
- An exploratory ASAP run with two selectivity blocks (1993-2011 and 2012-2015) estimated slightly higher lower selectivity for 1-year olds but slightly higher selectivity for 2 -year olds since the introduction of the panels. The assessment results were otherwise nearly identical.
- The XSA assessment (which does not have a fixed selectivity pattern) does not showclear reductions in F for younger ages relative to the older ages since 2012.
- A change in selectivity may also be detected from a change in mean weight-at-age for young fish (within an age class the smaller, lighter fish should escape). The average catch weight of 1-year olds has shown minor increases since the late 2000s (Figure 7.4.8). The catch weights of 2-year-olds increased between 2011 and 2018, while three-year-olds have also shown an increasing trend, and this age class is not expected to be affected by square-mesh panels.

Therefore there is no clear evidence that selectivity has changed significantly and the assumption of constant selectivity in ASAP appears to be valid. In future assessments a separate selectivity block for the last three years should continue to be considered.

### 9.8.4 Assessment bias

Figure 7.4.18 shows the retrospective of the ASAP analysis. The predicted catch shows little retrospective pattern, neither does the recruitment estimate. The SSB however, has a tendency to be revised upwards as another year of data is added. F has been overestimated and revised downwards since approximately 2008 in the assessment, caused by the strong 2009 cohort for which caused a conflict between the catch data, and being most pronounced in 2011 and 2012, becoming less with the addition of another year. This retrospective bias appears to become more aligned form 2013-2014 onwards.

Assessment bias (Figure 7.4.19) is more apparent for SSB in the most recent year, being revised up, while F and recruitment showed little revision historically, and only recruitment showing any marked change, in the 2018 value owing to the large recruitment observed in the surveys. Mohn's rho was calculated for the assessment, comparing the yearly estimates with proceeding years' assessment estimates, with comparisons lagged to remove the last year value from each assessment, as this value will have been influenced by recruitment, which was set to the geometric mean of the proceeding time-series values. The averages of the proceeding five years' Mohn's rho were calculated, for SSB $(-0.172) \mathrm{F}(0.113)$ and recruitment (0.003). In line with Figure 7.4.19, values for $F$ and recruitment were low. The value for SSB was strongly influenced by the upward shift in the most recent years' assessment, with proceeding years' assessments agreeing well. The increase in SSB observed in the most recent year's assessment is attributable to the observed 2018 large recruitment event, being well above recent and average level of recruitment in the stock. ICES recommendations for Mohn's rho is a range within 0.20 and -0.15 . The SSB value in this case falls outside this range, however owing to the bias being a response to an observed, unusual recruitment even it is considered that on this occasion the resulting advice is still fit for purpose.

### 9.9 Surveys

The combined French/Irish survey has nearly full spatial coverage of the assessment area. The survey has good internal consistency. The commercial tuning fleet only covers a small part of the stock area but WKROUND (2012) decided to include this fleet due to the short time-series of the survey.

### 9.9.1 Forecast

The 2018 cohort accounts for over $77 \%$ the projected landings in 2020, with recruitment of this cohort estimated with a CV of $21 \%$, which is reasonably precise and recruitment estimates have tended to be accurate in the past with little retrospective bias. The strong cohort was picked up by both the Irish and French quarter 4 surveys in 2018.

The 2018 GM recruitment assumption does not contribute much to the forecasted landings in 2020 ( $3 \%$ contribution), however it contributes $35 \%$ to the 2021 SSB estimate; this adds considerable uncertainty to the 2021 SSB forecast.

The large recruitment event seen in 2018 surveys is expected to contribute significantly to the realised 2019 catch and 2020 catch, (Figure 7.4.20, Table 7.4.11 \& 7.4.12).

### 9.10 Recommendation for next benchmark

### 9.10.1 Stock audit

The audit of the 2019 report did not raise any concerns.

### 9.10.2 Recommendations for future work

WGCSE recommend that cod, haddock and whiting in the Celtic Sea will be benchmarked together in 2019-2020. The focus of the benchmark will streamlining data compilation procedures for fishery-dependent and survey data. This will give improved transparency and diagnostics surrounding commercial tuning fleets and surveys. The benchmark will also re-examine the assessment methods and diagnostics given the potential for changes in selectivity in the commercial fishery. The benchmark should also consider mixed fisheries and multispecies interactions as well as environmental drivers that may be impacting on growth and recruitment of all three species.
Catch data should continue to be monitored for indirect evidence of improved selection patterns due to the augmented TCMs in the Celtic Sea. Direct monitoring of escapement through SMPs would also be useful.

It would be desirable to include discards separately in the assessment model in order to specify greater precision for the discard numbers-at-age than for the landings numbers-at-age. However WKROUND (2012) concluded that this resulted in undesirable residual patterns. The benchmark workshop did not have sufficient time to fully evaluate this problem.

It would be worth investigating the value of the commercial tuning fleet. If this fleet is to be retained it would be useful to apply a method of standardisation to account for possible changes in the fleet.

### 9.11 Management considerations

The stock size fluctuates strongly over the time. The size of the stock is determined to a large extent by recruitment, which has been erratic and in 2018 is shown to have been large. There is no discernible relationship between stock size and recruitment, as is the case with most haddock stocks.

Fishing mortality has been consistently above Fmsy, but this has not led to a decreasing trend in stock size, which suggests that the stock is robust to overfishing, however $F$ has been increasing since 2015 and at current levels the SSB could quickly fall below MS Ybtriger if recruitment were to be low for three or four years. Recent high 2018 recruitment is not yet appearing in the SSB estimates, which continues to fall.

The variable recruitment has also resulted in substantial short-term variability in TACs and high discards have occurred when a strong year class occurs, this is expected to be the case in 2019 and 2020 (Figure 7.4.19). Discarding of under-size as well as marketable fish is a serious problem for this stock, with approximately $2 / 3$ in catch numbers and almost half the catch weight has been discarded on average over the past decade. Alternative or complimentary approaches to managing such strong, recruit-driven fluctuations are required, especially with regard to the EU landings obligation.

The minimum landing size of haddock is 30 cm , which is approximately the same as the mean length of two-year old haddock in the Celtic sea. Because gadoids are caught in a mixed fishery,
restrictive quota in recent years have led to increased discarding of marketable fish as well as already considerable discarding of undersized fish. Technical measures have been introduced to reduce discards of undersize gadoids ( 110 mm square-mesh panel in the Nephrops fisheries and 100 mm in the gadoid fisheries). It is not clear whether this is sufficient to reduce discard mortality of future cohorts. It is important that technical measures are fully implemented and their effectiveness in reducing discards and impact on commercial catches are monitored and evaluated.

### 9.12 References

EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008.

ICES. 2016a. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMS YREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

ICES. 2016b. EU request to ICES to provide FMSY ranges for selected stocks in ICES subareas 5 to 10. ICES Advice 2016 Book 5, ICES Special Request Advice, Published 5 February 2016.

Table 7.4.1.a. Haddock in 7.bc-ek. Official landings (quota uptake in brackets).

| Year | BEL | ESP | FRA | IRL | UK* | Others | Total | TAC** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 123 | 0 | 2788 | 908 | 240 | 17 | 4076 |  |
| 1995 | 189 (28\%) | 19 | 2964 (74\%) | 966 (72\%) | 266 (44\%) | 64 | 4468 | 6000 |
| 1996 | 133 (9\%) | 48 | 4527 (49\%) | 1468 (47\%) | 439 (31\%) | 38 | 6653 | 14000 |
| 1997 | 246 (16\%) | 54 | 6581 (71\%) | 2789 (90\%) | 569 (41\%) | 31 | 10270 | 14000 |
| 1998 | 142 (6\%) | 260 | 3674 (28\%) | 2788 (63\%) | 445 (22\%) | 52 | 7361 | 20000 |
| 1999 | 51 (2\%) | 88 | 2725 (19\%) | 2034 (42\%) | 278 (13\%) | 71 | 5247 | 22000 |
| 2000 | 90 (5\%) | 110 | 3088 (28\%) | 3066 (83\%) | 289 (17\%) | 13 | 6656 | 16600 |
| 2001 | 165 (12\%) | 646 | 4842 (61\%) | 3608 (135\%) | 422 (35\%) | 19 | 9702 | 12000 |
| 2002 | 132 (128\%) |  | 4348 (70\%) | 2188 (106\%) | 315 (34\%) | 106 | 7089 | 9300 |
| 2003 | 118 (130\%) |  | 5781 (106\%) | 1867 (103\%) | 393 (48\%) | 82 | 8241 | 8185 |
| 2004 | 136 (127\%) |  | 6130 (96\%) | 1715 (80\%) | 313 (33\%) | 159 | 8453 | 9600 |
| 2005 | 167 (130\%) |  | 4166 (54\%) | 2037 (80\%) | 292 (25\%) | 197 | 6859 | 11520 |
| 2006 | 99 (77\%) |  | 3190 (42\%) | 1875 (73\%) | 274 (24\%) | 209 | 5647 | 11520 |
| 2007 | 119 (93\%) |  | 4142 (54\%) | 1930 (75\%) | 386 (34\%) | 52 | 6629 | 11520 |
| 2008 | 108 (84\%) |  | 3639 (47\%) | 1800 (70\%) | 566 (49\%) | 121 | 6234 | 11579 |
| 2009 | 131 (102\%) |  | 5429 (70\%) | 2983 (116\%) | 716 (62\%) | 48 | 9307 | 11579 |
| 2010 | 170 (132\%) |  | 6240 (81\%) | 2609 (101\%) | 852 (74\%) | 128 | 9999 | 11579 |
| 2011 | 211 (143\%) |  | 8388 (94\%) | 3322 (112\%) | 1659 (125\%) | 129 | 13709 | 13316 |
| 2012 | 231 (125\%) |  | 11793 (106\%) | 4130 (112\%) | 1901 (114\%) | 167 | 18222 | 16645 |
| 2013 | 173 (110\%) |  | 8748 (93\%) | 2699 (86\%) | 1455 (103\%) | 21 | 13096 | 14148 |
| 2014 | 99 (94\%) |  | 6374 (101\%) | 2092 (99\%) | 785 (83\%) | 18 | 9368 | 9479 |
| 2015 | 117 (126\%) |  | 5681 (102\%) | 1656 (89\%) | 759 (91\%) | 4 | 8217 | 8342 |
| 2016 | 88 (102\%) |  | 4487 (87\%) | 1713 (99\%) | 692 (89\%) | 27 | 7007 | 7751 |
| 2017 | 111 (144\%) |  | 4615 (100\%) | 1256 (82\%) | 690 (100\%) | 13 | 6685 | 6910 |
| 2018 | 89 (96\%) |  | 4478 (81\%) | 1434 (77\%) | 581 (70\%) | 8 | 6590 | 8329 |

* UK Includes Channel Islands.
** TAC Applied to subareas 7-10 from 1995 to 2008 and to $7 \mathrm{~b}-\mathrm{k}, 8,9$ and 10 from 2009 onwards.

Table 7.4.1.b. Haddock in 7.bc-ek. ICES estimate of the landings (lan) and discards (dis).

| Year | $\begin{aligned} & \text { BEL } \\ & \text { Lan } \end{aligned}$ | $\begin{aligned} & \text { ESP } \\ & \text { Lan } \end{aligned}$ | $\begin{aligned} & \text { FRA } \\ & \text { Lan } \end{aligned}$ | $\begin{aligned} & \text { IRL } \\ & \text { Lan } \end{aligned}$ | $\begin{aligned} & \text { UK } \\ & \text { Lan } \end{aligned}$ | Others Lan | Total Lan | $\begin{aligned} & \text { FRA } \\ & \text { Dis* } \end{aligned}$ | IRL Dis** | Others <br> Dis*** | Total Dis |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 |  |  |  |  |  |  | 3348 | 505 | 594 | 109 | 1208 |
| 1994 |  |  |  |  |  |  | 4131 | 1116 | 594 | 176 | 1886 |
| 1995 |  |  |  |  |  |  | 4470 | 730 | 1221 | 267 | 2218 |
| 1996 |  |  |  |  |  |  | 6756 | 3170 | 713 | 426 | 4309 |
| 1997 |  |  |  |  |  |  | 10827 | 2129 | 502 | 253 | 2883 |
| 1998 |  |  |  |  |  |  | 7928 | 680 | 140 | 114 | 934 |
| 1999 |  |  |  |  |  |  | 4970 | 477 | 54 | 55 | 586 |
| 2000 |  |  |  |  |  |  | 7499 | 1587 | 727 | 189 | 2503 |
| 2001 |  |  |  |  |  |  | 9278 | 2234 | 743 | 441 | 3418 |
| 2002 | 134 | 85 | 3878 | 2070 | 301 | 21 | 6488 | 871 | 5651 | 552 | 7073 |
| 2003 | 116 | 82 | 5960 | 1731 | 362 | 41 | 8292 | 1835 | 6941 | 680 | 9456 |
| 2004 | 137 | 143 | 6336 | 1785 | 303 | 73 | 8777 | 1108 | 5156 | 486 | 6750 |
| 2005 | 165 | 197 | 4096 | 2026 | 282 | 21 | 6787 | 762 | 3933 | 496 | 5191 |
| 2006 | 98 | 185 | 3151 | 1883 | 262 | 14 | 5593 | 1061 | 1167 | 256 | 2484 |
| 2007 | 118 | 49 | 4073 | 2135 | 383 | 23 | 6781 | 1268 | 1241 | 230 | 2739 |
| 2008 | 109 | 121 | 4587 | 2032 | 545 | 61 | 7455 | 7608 | 2153 | 1427 | 11187 |
| 2009 | 131 | 47 | 5455 | 3271 | 703 | 1 | 9608 | 6064 | 2143 | 873 | 9080 |
| 2010 | 170 | 127 | 6267 | 2876 | 789 | 34 | 10262 | 11396 | 3246 | 1905 | 16547 |
| 2011 | 212 | 94 | 7365 | 3697 | 1511 | 0 | 12879 | 9320 | 2913 | 2145 | 14378 |
| 2012 | 232 | 105 | 11793 | 4608 | 1637 | 0 | 18376 | 7221 | 1678 | 1293 | 10191 |
| 2013 | 174 | 40 | 8622 | 3109 | 1480 | 0 | 13424 | 1103 | 727 | 255 | 2085 |
| 2014 | 99 | 3 | 6376 | 2529 | 848 | 0 | 9855 | 1793 | 992 | 392 | 3177 |
| 2015 | 118 | 0 | 5679 | 1978 | 766 | 4 | 8545 | 2798 | 2785 | 1110 | 6693 |
| 2016 | 88 | 0 | 4487 | 1713 | 692 | 26 | 7574 |  |  |  |  |
| 2017 | 111 | 0.180 | 4896 | 2379 | 699 | 0 | 8086 | 4357 | 1597 | 2021 | 7975 |
| 2018 | 89 | 0 | 4446 | 1986 | 578 | 7 | 7109 | 2733 | 1133 | 1570 | 5436 |

* For 1993-2007 fixed discard ratios were used to estimate French discards.
** For 1993-1994, the mean Irish discards over 1995-1999 were used.
*** Estimated from the proportion of the landings of `Others' between 1993 and 2012.

Table 7.4.2. Haddock in 7.bc-ek. Lpue (kg/hour fishing) of haddock and effort (hours fishing x 1000) for Irish Otter trawls in 7.bc, 7.fgh and 7.jk, the French demersal fleet in 7.bc-ek and effort only for the UK trawl fleets (excluding beam trawls) in 7.e-k (effort in fishing days).

|  | FR GAD <br> 7ek effort | FR GAD <br> 7ek lpue | IRL OTB <br> 7bc effort | IRL OTB <br> 7bc lpue | IRL OTB 7fgh effort | IRL OTB 7fgh Ipue | IRL OTB <br> 7jk effort | IRL OTB <br> 7jk Ipue | UK Trawl 7e-k effort |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | NA | NA | NA | NA | NA | NA | NA | NA | 51.5 |
| 1984 | NA | NA | NA | NA | NA | NA | NA | NA | 161.8 |
| 1985 | NA | NA | NA | NA | NA | NA | NA | NA | 143.7 |
| 1986 | NA | NA | NA | NA | NA | NA | NA | NA | 123.5 |
| 1987 | NA | NA | NA | NA | NA | NA | NA | NA | 108.9 |
| 1988 | NA | NA | NA | NA | NA | NA | NA | NA | 112.9 |
| 1989 | NA | NA | NA | NA | NA | NA | NA | NA | 119.9 |
| 1990 | NA | NA | NA | NA | NA | NA | NA | NA | 133.2 |
| 1991 | NA | NA | NA | NA | NA | NA | NA | NA | 118.8 |
| 1992 | NA | NA | NA | NA | NA | NA | NA | NA | 129.9 |
| 1993 | NA | NA | NA | NA | NA | NA | NA | NA | 101.1 |
| 1994 | NA | NA | NA | NA | NA | NA | NA | NA | 88.5 |
| 1995 | NA | NA | 78 | 5.77 | 64 | 1.48 | 106 | 2.20 | 88.1 |
| 1996 | NA | NA | 47 | 4.16 | 60 | 5.35 | 73 | 3.24 | 89.5 |
| 1997 | NA | NA | 63 | 4.36 | 65 | 5.83 | 92 | 8.23 | 101.8 |
| 1998 | NA | NA | 79 | 5.71 | 72 | 4.09 | 99 | 5.88 | 94.6 |
| 1999 | NA | NA | 77 | 5.27 | 51 | 2.35 | 52 | 3.53 | 132.8 |
| 2000 | 306 | 6.12 | 74 | 4.73 | 61 | 10.43 | 72 | 4.25 | 141.1 |
| 2001 | 333 | 10.57 | 78 | 4.30 | 69 | 8.69 | 81 | 7.41 | 117.5 |
| 2002 | 289 | 10.63 | 63 | 2.81 | 79 | 3.22 | 108 | 5.50 | 113.1 |
| 2003 | 264 | 15.15 | 81 | 2.09 | 87 | 3.26 | 123 | 3.88 | 102.4 |
| 2004 | 217 | 19.39 | 82 | 2.51 | 97 | 3.49 | 108 | 3.35 | 105.5 |
| 2005 | 175 | 14.67 | 69 | 2.45 | 127 | 4.53 | 93 | 3.70 | 100.9 |
| 2006 | 167 | 10.64 | 60 | 2.56 | 119 | 4.19 | 89 | 3.59 | 106.3 |
| 2007 | 160 | 14.97 | 60 | 3.31 | 136 | 4.01 | 103 | 3.66 | 113.6 |
| 2008 | 148 | 19.60 | 48 | 4.36 | 127 | 4.56 | 84 | 4.60 | 93.7 |
| 2009 | 150 | 22.65 | 48 | 5.47 | 141 | 9.25 | 82 | 7.09 | 98.6 |
| 2010 | 131 | 30.83 | 54 | 4.36 | 144 | 7.33 | 101 | 5.15 | 103.7 |
| 2011 | 216 | 22.90 | 40 | 6.39 | 129 | 10.51 | 84 | 5.58 | 87.1 |
| 2012 | 188 | 45.03 | 44 | 4.93 | 135 | 13.17 | 84 | 6.58 | 86.2 |
| 2013 | 215 | 27.40 | 42 | 5.38 | 126 | 8.69 | 80 | 4.92 | 40.3 |
| 2014 | 203 | 19.81 | 46 | 5.22 | 142 | 5.11 | 77 | 3.91 | 32.1 |
| 2015 | NA | NA | 31 | 4.42 | 150 | 4.95 | 78 | 2.91 | 21.2 |
| 2016 | NA | NA | 39 | 2.41 | 164 | 4.94 | 83 | 3.09 | NA |
| 2017 | NA | NA | 36 | 2.25 | 151 | 5.10 | 92 | 2.43 | NA |
| 2018 | NA | NA | 46 | 2.19 | 125 | 5.33 | 93 | 1.70 | NA |

Table 7.4.3a. Haddock in 7.bc-ek. Landings numbers-at-age.

|  | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 491 | 3291 | 948 | 810 | 255 | 129 | 129 | 45 |
| 1994 | 0 | 1277 | 5223 | 674 | 302 | 94 | 24 | 35 | 16 |
| 1995 | 0 | 4275 | 1622 | 1327 | 270 | 245 | 46 | 0 | 0 |
| 1996 | 0 | 3693 | 15998 | 818 | 313 | 93 | 32 | 10 | 9 |
| 1997 | 0 | 1353 | 9645 | 5553 | 716 | 354 | 139 | 144 | 110 |
| 1998 | 0 | 167 | 3184 | 7403 | 1443 | 307 | 178 | 86 | 61 |
| 1999 | 0 | 476 | 654 | 1464 | 2425 | 307 | 18 | 19 | 6 |
| 2000 | 0 | 2197 | 2996 | 784 | 741 | 1250 | 205 | 35 | 28 |
| 2001 | 0 | 4297 | 8638 | 1131 | 303 | 317 | 321 | 54 | 39 |
| 2002 | 0 | 879 | 4274 | 3400 | 765 | 39 | 89 | 74 | 26 |
| 2003 | 0 | 703 | 8791 | 2160 | 1226 | 116 | 43 | 49 | 51 |
| 2004 | 0 | 125 | 5948 | 4663 | 928 | 589 | 51 | 12 | 20 |
| 2005 | 0 | 786 | 863 | 4366 | 1983 | 450 | 115 | 4 | 17 |
| 2006 | 0 | 852 | 3393 | 1500 | 2219 | 400 | 67 | 7 | 1 |
| 2007 | 0 | 707 | 6404 | 2687 | 532 | 864 | 155 | 29 | 5 |
| 2008 | 0 | 1637 | 4034 | 4422 | 987 | 235 | 382 | 70 | 13 |
| 2009 | 0 | 795 | 7010 | 3394 | 1939 | 489 | 145 | 110 | 27 |
| 2010 | 0 | 1291 | 4814 | 6091 | 901 | 494 | 162 | 68 | 62 |
| 2011 | 0 | 170 | 11164 | 3359 | 3249 | 606 | 200 | 55 | 43 |
| 2012 | 0 | 61 | 787 | 18587 | 2352 | 1319 | 212 | 60 | 54 |
| 2013 | 0 | 24 | 244 | 2071 | 11007 | 764 | 444 | 87 | 47 |
| 2014 | 0 | 284 | 719 | 309 | 1632 | 5587 | 272 | 108 | 19 |
| 2015 | 0 | 111 | 4775 | 552 | 215 | 946 | 1896 | 165 | 23 |
| 2016 | 0 | 58 | 312 | 5543 | 202 | 95 | 402 | 476 | 45 |
| 2017 | 0 | 21 | 1808 | 960 | 3874 | 113 | 57 | 191 | 203 |
| 2018 | 0 | 156 | 715 | 3347 | 728 | 1435 | 40 | 38 | 102 |

Table 7.4.3b. Haddock in 7.bc-ek. Discard numbers-at-age.

|  | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 7617 | 2816 | 160 | 6 | 0 | 0 | 0 | 0 |
| 1994 | 0 | 15120 | 3069 | 170 | 5 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 32830 | 1977 | 91 | 4 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 20734 | 8976 | 187 | 9 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 12613 | 10022 | 493 | 5 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 3580 | 2348 | 445 | 5 | 0 | 0 | 0 | 0 |
| 1999 | 0 | 3742 | 1562 | 100 | 10 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 29015 | 2521 | 64 | 3 | 0 | 0 | 0 | 0 |
| 2001 | 0 | 25234 | 6772 | 219 | 2 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 21624 | 20729 | 249 | 7 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 52412 | 11075 | 352 | 8 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 11733 | 21598 | 1395 | 61 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 15904 | 10766 | 4315 | 149 | 0 | 0 | 0 | 0 |
| 2006 | 0 | 9377 | 4130 | 381 | 33 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 6387 | 7066 | 662 | 34 | 0 | 0 | 0 | 0 |
| 2008 | 0 | 48764 | 15658 | 5492 | 330 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 23561 | 27015 | 873 | 581 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 98400 | 23292 | 2133 | 131 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 16081 | 47971 | 1831 | 665 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 7056 | 22315 | 12250 | 115 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 1645 | 1187 | 1339 | 1899 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 13089 | 3385 | 449 | 176 | 155 | 0 | 0 | 0 |
| 2015 | 0 | 2806 | 17841 | 550 | 14 | 103 | 134 | 15 | 1 |
| 2016 | 0 | 22590 | 4116 | 6993 | 80 | 4 | 33 | 311 | 0 |
| 2017 | 0 | 4389 | 12077 | 1279 | 2268 | 40 | 3 | 12 | 6 |
| 2018 | 0 | 4487 | 4277 | 4073 | 384 | 318 | 3 | 2 | 5 |

Table 7.4.4. Haddock in 7.bc-ek. ASAP input data.

```
# ASAP VERSION 3.0
# Had7b-k
#
# ASAP GUI 15 AUG 2012
#
# Number of Years
26
# First Year
1993
# Number of Ages
9
# Number of Fleets
1
# Number of Sensitivity Blocks
1
# Number of Available Survey Indices
```



```
# Natural Mortality
0.99 0.72 0.6 0.5 0.43 0.4 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.5 0.43 0.4 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99}0.720.720.6 0.5 0.43 0.4 0.47 0.3 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.47 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.5 0.43 0.4 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.4 0.43 0.4 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.4 0.43 0.4 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
0.99 0.72 0.6 0.5 0.43 0.4 0.37 0.36 0.34
# Fecundity Option
0
# Fraction of year that elapses prior to SSB calculation (0=Jan-1)
0
# Maturity
0 0 1 1 1 1 1 1 1
0 0 1 1 1 1 1 1 1 1
0 0 1 1 1 1 1 1 1 1
0 0 1 1 1 1 1 1 1
```



```
0 0 1 1 1 1 1 1 1
0}001111111111 1 1
0 0 1 1 1 1 1 1 1 1
```




```
0 0 1 1 1 1 1 1 1
0}00111111111111
0 0 1 1 1 1 1 1 1
```



```
0 0 1 1 1 1 1 1 1
```




```
0 0 1 1 1 1 1 1 1
0 0 1 1 1 1 1 1 1
```



```
0 0 1 1 1 1 1 1 1
0011111111
```

```
0}00111111 1 1 1 1
0}01111111111
0 0 1 1 1 1 1 1 1
0 0 1 1 1 1 1 1 1
# Number of Weights at Age Matrices
2
# Weight Matrix - 1
0 0.09 0.257 0.524 0.848 1.402 1.693 2.13 2.573
0 0.1 0.358 0.614 0.987 1.456 1.745 2.014 2.536
0 0.089 0.388 0.875 1.321 1.188 1.746 0 0
0}0.130.275 0.576 0.799 1.181 1.369 1.828 1.827
0 0.097 0.305 0.743 1.205 1.362 1.268 1.412 1.176
0 0.103 0.296 0.611 0.938 0.956 1.086 1.292 1.453
0 0.129 0.299 0.848 1.072 1.186 1.223 0.908 1.708
0 0.091 0.452 1.19 1.463 1.719 1.627 1.163 1.459
0 0.122 0.384 0.971 1.857 1.783 1.705 2.297 1.612
0 0.095 0.295 0.791 1.03 1.733 1.678 1.505 1.569
0}00.133 0.353 0.804 1.238 1.441 1.818 1.704 1.709
0 0.136 0.285 0.654 1.135 1.378 1.876 1.84 2.084
0}0.1360.211 0.499 0.971 1.252 1.942 2.667 1.949
0 0.162 0.348 0.504 0.925 1.47 2.091 2.59 4.022
0 0.168 0.34 0.566 0.855 1.2 1.642 1.507 2.837
0 0.13 0.287 0.461 0.74 1.159 1.282 1.685 1.926
0 0.118 0.291 0.618 0.846 1.311 1.547 1.653 2.441
0 0.114 0.268 0.653 1.072 1.754 1.845 1.738 1.673
0 0.155 0.278 0.59 0.928 1.623 2.116 1.888 1.478
0 0.127 0.248 0.543 1.041 1.443 2.022 2.278 2.203
0}00.151 0.298 0.587 0.832 1.422 1.611 2.209 1.86
0 0.142 0.372 0.63 0.911 1.179 1.654 1.965 2.576
0 0.155 0.403 0.667 1.02 1.233 1.478 1.859 2.462
0 0.2 0.3 0.7 1.022 1.448 1.709 1.566 2.017
0 0.214 0.41 0.673 1.086 1.318 1.998 2.013 2.023
0 0.267 0.42 0.692 1.056 1.464 1.673 1.765 2.026
# Weight Matrix - 2
0.041 0.093 0.277 0.641 0.824 1.804 2.089 2.407 2.647
0.042 0.093 0.29 0.756 1.138 2.36 2.163 2.407 2.647
0.045 0.102 0.295 0.715 1.232 2.174 1.972 2.169 2.386
0.046 0.1 0.313 0.719 1.246 2.046 1.773 1.95 2.145
0.043 0.098 0.287 0.579 0.904 1.144 1.261 1.631 1.794
0.037 0.096 0.274 0.655 0.87 1.005 1.016 1.251 1.376
0.028 0.103 0.265 0.791 0.962 1.148 1.203 1.348 1.483
0.027 0.109 0.306 0.93 1.326 1.548 1.605 1.765 1.942
0.022 0.102 0.312 0.926 1.33 1.634 1.672 1.84 2.024
0.021 0.11 0.312 0.841 1.399 1.676 1.888 2.076 2.284
0.023 0.119 0.275 0.725 1.189 1.601 1.938 2.132 2.345
0.032 0.133 0.248 0.623 1.207 1.662 2.308 2.538 2.792
0.037 0.139 0.252 0.523 1.056 1.587 2.159 2.409 2.65
0.043 0.148 0.265 0.49 0.922 1.417 2.062 2.537 2.79
0.041 0.145 0.282 0.481 0.799 1.313 1.763 2.168 2.385
0.048 0.135 0.267 0.505 0.759 1.148 1.611 1.838 2.022
0.048 0.119 0.252 0.522 0.804 1.252 1.519 1.775 1.952
0.041 0.128 0.256 0.55 0.861 1.331 1.732 2.036 2.24
0.043 0.13 0.251 0.52 0.913 1.439 1.896 2.268 2.495
0.044 0.142 0.263 0.512 0.87 1.445 1.95 2.514 2.765
0.054 0.137 0.291 0.55 0.886 1.407 1.867 2.384 2.622
0.055 0.147 0.333 0.604 0.894 1.4 1.695 2.167 2.384
0.064 0.164 0.336 0.642 0.957 1.407 1.727 1.924 2.116
0.063 0.189 0.351 0.656 1.016 1.45 1.848 2.033 2.236
0.061 0.225 0.36 0.665 1.03 1.53 1.918 2.11 2.321
0.05 0.238 0.394 0.659 1.047 1.503 1.964 2.16 2.376
# Weights at Age Pointers
1
1
1
1
2
# Selectivity Block Assignment
# Fleet 1 Selectivity Block Assignment
1
1
1
1
```

```
\rightharpoonup
1
1
1
1
1
1
1
1
1
1
1
1
1
1
# Selectivity Options for each block 1=by age, 2=logisitic, 3=double logistic
1
# Selectivity Block #1 Data
0-1 0 1
0.5 1 0 1
1 1 0 1
1 -1 0 1
1 -1 0 1
1 -1 0 1
1 -1 0 1
1 -1 0 1
1 -1 0 1
1101
1101
1101
1101
1101
1101
# Fleet Start Age
1
# Fleet End Age
9
# Age Range for Average F
46
# Average F report option (1=unweighted, 2=Nweighted, 3=Bweighted)
1
# Use Likelihood constants? (1=yes)
0
# Release Mortality by Fleet
1
# Catch Data
# Fleet-1 Catch Data
0 8107 6107 1108 816 255 129 129 45 4556
0 16396 8292 844 307 94 24 35 16 6017
0 37105 3599 1419 273 24546 0 0 6688
0 24428 24973 1005 321 93 32 10 9 11065
0 13965 19667 6046 722 354 139 144 110 13710
0 3747 55317848 1448 307 178 86 61 8862
0 4218 2217 1564 2435 307 18 19 6 5556
0 31212 5517 848 744 1250 205 35 28 10002
0 29531 15409 1350 304 317 321 54 39 12696
0 22503 25003 3650 772 39 89 74 26 13561
0 53115 19866 2512 1234 116 43 49 51 17748
0 11858 27546 6058 989 589 51 12 20 15527
0 16690 11629 8681 2133 450 115 4 17 11978
0 10229 7524 1881 225240067 7 1 8077
0 7094 13470 3350 566 864 155 29 5 9520
0 50401 19692 9913 1317 235 382 70 13 18642
0
0 99691 28106 8225 1033 494 162 68 62 26809
0 1625259134 5190 3914 606 200 55 43 27257
0 7116 23102 30837 2467 1319 212 60 54 28567
0 1669 1431 341012906 764 444 87 47 15509
```

```
0 13372 4103 758 1808 5741 272 108 19 13031
0 2918 22616 1102 229 1049 2029 180 24 15239
0 22477 4503 13233 312 100 443 803 46 17931
0 4410 13885 2239 6141 153 60 203 211 16072
0 4644 4992 7420 1112 1753 43 41 106 12546
# Discards
# Fleet-1 Discards Data
0 0 0 0 0 0 0 0 0 0
000000000000
00000000000
00000000000
00000000000
00000000000
0 0 0 0 0 0 0 0 0 0
00000000000
0000000000
00 0 0 0 0 0 0 0 0
000000000000
00000000000
00000000000
0 0 0 0 0 0 0 0 0 0
00000000000
00000000000
0 0 0 0 0 0 0 0 0 0
000000000000
00000000000
0 0 0 0 0 0 0 0 0 0
00000000000
00000000000
00000000000
00000000000
00000000000
00000000000
# Release Proportion
# Fleet-1 Release Data
0 0 0 0 0 0 0 0 0
0000000000
0 0 0 0 0 0 0 0 0
0000000000
0000000000
000000000
0000000000
000000000
000000000
000000000
0000000000
00 0 0 0 0 0 0 0
0000000000
000000000
00 0 0 0 0 0 0 0
0000000000
0 0 0 0 0 0 0 0 0
000000000
0 0 0 0 0 0 0 0 0
000000000
0000000000
0000000000
000000000
00 0 0 0 0 0 0 0
0000000000
000000000
# Survey Index Data
# Aggregate Index Units
2 2
# Age Proportion Index Units
2 2
# Weight at Age Matrix
2 2
# Index Month
117
# Index Selectivity Link to Fleet
-1 -1
# Index Selectivity Options 1=by age, 2=logisitic, 3=double logistic
11
```

```
# Index Start Age
14
# Index End Age
6 8
# Estimate Proportion (Yes=1)
1 1
# Use Index (Yes=1)
1 1
# Index-1 Selectivity Data
1 1 1 1e-04
1-1 0 1
1 -1 0 1
1 -1 0 1
1 -1 0 1
1-1 0 1
-1 -1 0 1
-1 -1 0 1
-1 -1 0 1
1101
1101
0-1 0 1
0.001 -1 0 1
1101
1101
# Index-2 Selectivity Data
-1 -1 0 1
-1 -1 0 1
-1 -1 0 1
0.8 1 0 1
1 -1 0 1
1-1 0 1
1 -1 0 1
1 -1 0 1
-1 -1 0 1
1101
1101
3-1 0 1
1-1 0 1
8-1 0 1
1 -1 0 1
# Index-1 Data
199300000000000000
199400000 0 0 0 0 0 0 0 0
19950000000000000
199600000000000000
199700000000000000
199800000000000000
199900000000000 0 0 0
200000000000000000
2001 0 0 0 0 0 0 0 0 0 0 0 0
200200000000000000
2003 707.4 0.2 157 508.3 32.67 2.4 0.1 0 0 0 40
2004 517.7 0.2 385.7 49.1 70.9 7.9 2.7 1.4 0 0 0 40
2005 310.7 0.2 193.5 85.7 9.9 19.4 1.9 0.3 0 0 0 40
2006 176.9 0.2 110.2 39.7 19 4.5 3.2 0.4 0 0 0 40
2007 670.6 0.2 610.8 38.6 9.9 5.8 2.8 2.7 0 0 0 40
2008 424 0.2 271.5 143.3 5.6 1.6 1.3 0.7 0 0 0 40
2009 1562.4 0.2 1428.4 67.1 62 2.1 1.9 0.8 0 0 0 40
2010 823.4 0.2 89.7 686 33 13.6 0.4 0.8 0 0 0 40
2011 317.8 0.2 69.2 45.3 193.9 7.2 2.1 0. 2 0 0 0 40
2012 113.9 0.2 21.4 23.1 13.4 52.4 2.2 1.3 0 0 0 40
2013 705.9 0.2 666 10.5 8.9 5.2 14.3 0.8 0 0 0 40
2014 279.9 0.2 91.3 177.2 2.4 1.9 2.1 5.1 0 0 0 40
2015 476.7 0.2 355.6 74.1 42.7 0.9 1.2 2.2 0 0 0 40
2016 248.6 0.2 43.1 155.7 39.3 9.3 0.9 0.3 0 0 0 40
2017 146.6 0.2 71.4 25.4 40.5 6.7 2.4 0.2 0 0 0 40
2018 733.8 0.2 696.5 16.5 4.9 7.2 6.5 2.2 0 0 0 40
# Index-2 Data
19930 0 0 0 0 0 0 0 0 0 0 0
19940000000000000000
19950.826 0.300000.751 0.06 0.015 0 0 0 40
1996 1.031 0.3 0 0 0 0.675 0.226 0.096 0.035 0 0 40
1997 3.578 0.3 0 0 0 3.086 0.339 0.115 0.019 0.019 0 40
1998 6.695 0.3 0 0 0 5.811 0.824 0.033 0.008 0.018 0 40
```

```
1999 3.047 0.3 0 0 0 1.147 1.735 0.149 0.005 0.011 0 40
2000 4.103 0.3 0 0 0 1.618 1.077 1.204 0. 204 0 0 40
2001 3.47 0.3 0 0 0 2.926 0.293 0.148 0.093 0.009 0 40
2002 3.996 0.3 0 0 0 3.657 0.266 0.02 0.021 0.034 0 40
2003 2.075 0.3 0 0 0 1.267 0.703 0.082 0.009 0.015 0 40
2004 4.594 0.3 0 0 0 3.368 0.858 0.351 0.01 0.008 0 40
2005 7.108 0.3 0 0 0 4.707 2.085 0.268 0.048 0 0 40
2006 7.058 0.3 0 0 0 2.976 3.523 0.484 0.062 0.012 0 40
2007 4.706 0.3 0 0 0 2.664 0.674 1.219 0.136 0.012 0 40
2008 5.48 0.3 0 0 0 3.56 1.17 0.258 0.404 0.088 0 40
2009 5.872 0.3 0 0 0 2.952 1.822 0.569 0.307 0.223 0 40
2010 9.978 0.3 0 0 0 8.297 0.964 0.506 0.154 0.057 0 40
2011 9.597 0.3 0 0 0 3.939 4.592 0.705 0.301 0.06 0 40
2012 17.739 0.3 0 0 0 13.829 1.746 1.787 0.285 0.092 0 40
2013 9.851 0.3 0 0 0 0.796 7.03 0.989 0.891 0.145 0 40
2014 4.997 0.3 0 0 0 0.225 0.972 3.584 0.155 0.061 0 40
2015 3.055 0.3 0 0 0 0.378 0.166 0.521 1.901 0.089 0 40
2016 5.165 0.3 0 0 0 4.305 0.126 0.049 0.254 0.431 0 40
2017 3.721 0.3 0 0 0 0.953 2.586 0.069 0.031 0.082 0 40
2018 5.614 0.3 0 0 0 3.383 0.752 1.377 0.064 0.037 0 0
# Phase Control
# Phase for F mult in 1st Year
1
# Phase for F mult Deviations
2
# Phase for Recruitment Deviations
3
# Phase for N in 1st Year
1
# Phase for Catchability in 1st Year
3
# Phase for Catchability Deviations
-5
# Phase for Stock Recruitment Relationship
1
# Phase for Steepness
-5
# Recruitment CV by Year
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
# Lambdas by Index
1 1
# Lambda for Total Catch in Weight by Fleet
1
# Lambda for Total Discards at Age by Fleet
1
# Catch Total CV by Year and Fleet
0.2
0.3
0.3
```

```
0.3
0.3
0.3
0.3
0.3
0.3
0.3
0.3
0.3
0.3
0.2
0.2
0.2
0.2
0.2
0.2
0.2
0.2
0.2
# Discard Total CV by Year and Fleet
0
0
0
0
0
0
0
0
0
0
0
0
# Catch Effective Sample Size by Year and Fleet
25
25
25
25
25
25
50
50
50
50
50
50
50
50
50
```

```
50
50
# Discard Effective Sample Size by Year and Fleet
0
0
0
0
0
0
0
0
0
0
0
0
0
0
0
0
0
# Lambda for F Mult in First year by Fleet
0
# CV for F Mult in First year by Fleet
0.5
# Lambda for F Mult Deviations by Fleet
0
# CV for F Mult Deviations by Fleet
0.5
# Lambda for N in 1st Year Deviations
0
# CV for N in 1st Year Deviations
1
# Lambda for Recruitment Deviations
0
# Lambda for Catchability in First year by Index
0
# CV for Catchability in First year by Index
1 1
# Lambda for Catchability Deviations by Index
0
# CV for Catchability Deviations by Index
1
# Lambda for Deviation from Initial Steepness
# CV for Deviation from Initial Steepness
1
# Lambda for Deviation from Unexploited Stock Size
0
# CV for Deviation from Unexploited Stock Size
1
# NAA Deviations Flag
1
# Initial Numbers at Age in 1st Year
40000 20000 10000 4000 2000 1000 500 250 100
# Initial F Mult in 1st Year by Fleet
0 . 7
# Initial Catchabilty by Index
1 1
# Stock Recruitment Flag
0
# Initial Unexploited Stock
1000
# Initial Steepness
```

```
# Maximum F
2.5
# Ignore Guesses (Yes=1)
0
# Projection Control
# Do Projections (Yes=1)
0
# Fleet Directed Flag
1
# Final Year in Projection
2019
# Projection Data by Year
2019 -1 3 -99 1
# Do MCMC (Yes=1)
0
# MCMC Year Option
0
# MCMC Iterations
1000
# MCMC Thinning Factor
200
# MCMC Random Seed
1415963
# Agepro R Option
0
# Agepro R Option Start Year
1993
# Agepro R Option End Year
2005
# Export R Flag
1
# Test Value
-23456
######
###### FINIS ######
# Fleet Names
#$LAND+DIS
# Survey Names
#$FR-IRL-IBTS
#$IR-GAD
#
```

Table 7.4.5. Haddock in 7.bc-ek. Selectivity of the catches and indices. Catch selectivity was fixed at zero for age 0 and at one for ages $3-8$; it was freely estimated for ages 1-2. For the FR_IR_IBTS survey the selectivity was fixed at 1 for all ages and for the IR_GAD commercial fleet selectivity was freely estimated for age 3 and fixed at 1 for the older ages. Catch and index selectivity were not allowed to vary over time.

| Age | Catch | FRA.IRL.IBTS | IRL.GAD |
| :---: | :---: | :---: | :---: |
| 0 | 0.000 | 1 | NA |
| 1 | 0.364 | 1 | NA |
| 2 | 0.980 | 1 | NA |
| 3 | 1.000 | 1 | 0.781 |
| 4 | 1.000 | 1 | 1.000 |
| 5 | 1.000 | 1 | 1.000 |
| 6 | 1.000 | NA | 1.000 |
| 7 | 1.000 | NA | 1.000 |
| 8 | 1.000 | NA | NA |

Table 7.4.6. Haddock in 7.bc-ek. Fishing mortality- (F) at-age.

|  | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 0.398 | 1.037 | 1.094 | 1.094 | 1.094 | 1.094 | 1.094 | 1.094 |
| 1994 | 0 | 0.388 | 1.011 | 1.067 | 1.067 | 1.067 | 1.067 | 1.067 | 1.067 |
| 1995 | 0 | 0.306 | 0.797 | 0.841 | 0.841 | 0.841 | 0.841 | 0.841 | 0.841 |
| 1996 | 0 | 0.303 | 0.790 | 0.833 | 0.833 | 0.833 | 0.833 | 0.833 | 0.833 |
| 1997 | 0 | 0.256 | 0.667 | 0.704 | 0.704 | 0.704 | 0.704 | 0.704 | 0.704 |
| 1998 | 0 | 0.275 | 0.715 | 0.754 | 0.754 | 0.754 | 0.754 | 0.754 | 0.754 |
| 1999 | 0 | 0.189 | 0.492 | 0.519 | 0.519 | 0.519 | 0.519 | 0.519 | 0.519 |
| 2000 | 0 | 0.240 | 0.624 | 0.658 | 0.658 | 0.658 | 0.658 | 0.658 | 0.658 |
| 2001 | 0 | 0.254 | 0.661 | 0.697 | 0.697 | 0.697 | 0.697 | 0.697 | 0.697 |
| 2002 | 0 | 0.455 | 1.184 | 1.249 | 1.249 | 1.249 | 1.249 | 1.249 | 1.249 |
| 2003 | 0 | 0.237 | 0.618 | 0.652 | 0.652 | 0.652 | 0.652 | 0.652 | 0.652 |
| 2004 | 0 | 0.283 | 0.736 | 0.777 | 0.777 | 0.777 | 0.777 | 0.777 | 0.777 |
| 2005 | 0 | 0.299 | 0.777 | 0.820 | 0.820 | 0.820 | 0.820 | 0.820 | 0.820 |
| 2006 | 0 | 0.187 | 0.487 | 0.513 | 0.513 | 0.513 | 0.513 | 0.513 | 0.513 |
| 2007 | 0 | 0.149 | 0.388 | 0.409 | 0.409 | 0.409 | 0.409 | 0.409 | 0.409 |
| 2008 | 0 | 0.265 | 0.689 | 0.727 | 0.727 | 0.727 | 0.727 | 0.727 | 0.727 |
| 2009 | 0 | 0.208 | 0.542 | 0.572 | 0.572 | 0.572 | 0.572 | 0.572 | 0.572 |
| 2010 | 0 | 0.216 | 0.562 | 0.593 | 0.593 | 0.593 | 0.593 | 0.593 | 0.593 |
| 2011 | 0 | 0.169 | 0.439 | 0.463 | 0.463 | 0.463 | 0.463 | 0.463 | 0.463 |
| 2012 | 0 | 0.209 | 0.544 | 0.574 | 0.574 | 0.574 | 0.574 | 0.574 | 0.574 |
| 2013 | 0 | 0.180 | 0.468 | 0.494 | 0.494 | 0.494 | 0.494 | 0.494 | 0.494 |
| 2014 | 0 | 0.184 | 0.478 | 0.504 | 0.504 | 0.504 | 0.504 | 0.504 | 0.504 |
| 2015 | 0 | 0.167 | 0.435 | 0.459 | 0.459 | 0.459 | 0.459 | 0.459 | 0.459 |
| 2016 | 0 | 0.211 | 0.550 | 0.581 | 0.581 | 0.581 | 0.581 | 0.581 | 0.581 |
| 2017 | 0 | 0.222 | 0.577 | 0.609 | 0.609 | 0.609 | 0.609 | 0.609 | 0.609 |
| 2018 | 0 | 0.275 | 0.715 | 0.755 | 0.755 | 0.755 | 0.755 | 0.755 | 0.755 |

Table 7.4.7. Haddock in 7.bc-ek. Stock numbers-at-age (start of year) (`1000).

|  | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 112280 | 50443 | 11948 | 2819 | 801 | 254 | 254 | 223 | 74 |
| 1994 | 384344 | 41721 | 16486 | 2324 | 572 | 174 | 57 | 59 | 70 |
| 1995 | 534466 | 142813 | 13773 | 3292 | 485 | 128 | 40 | 14 | 31 |
| 1996 | 151443 | 198595 | 51188 | 3407 | 861 | 136 | 37 | 12 | 14 |
| 1997 | 77087 | 56273 | 71376 | 12754 | 898 | 244 | 40 | 11 | 8 |
| 1998 | 160867 | 28644 | 21198 | 20098 | 3825 | 289 | 81 | 14 | 7 |
| 1999 | 424334 | 59775 | 10595 | 5692 | 5733 | 1170 | 91 | 26 | 7 |
| 2000 | 400838 | 157672 | 24084 | 3554 | 2054 | 2219 | 467 | 37 | 14 |
| 2001 | 456207 | 148942 | 60395 | 7083 | 1116 | 692 | 770 | 167 | 19 |
| 2002 | 803719 | 169516 | 56246 | 17116 | 2139 | 362 | 231 | 265 | 65 |
| 2003 | 220145 | 298643 | 52370 | 9451 | 2977 | 399 | 69 | 46 | 66 |
| 2004 | 284915 | 81801 | 114638 | 15488 | 2985 | 1009 | 139 | 25 | 41 |
| 2005 | 274362 | 105868 | 30010 | 30132 | 4320 | 893 | 311 | 44 | 21 |
| 2006 | 202585 | 101946 | 38230 | 7570 | 8047 | 1237 | 264 | 95 | 20 |
| 2007 | 712753 | 75276 | 41164 | 12898 | 2748 | 3133 | 496 | 109 | 48 |
| 2008 | 371945 | 264842 | 31573 | 15333 | 5197 | 1187 | 1395 | 228 | 73 |
| 2009 | 1755911 | 138206 | 98929 | 8697 | 4494 | 1634 | 385 | 466 | 102 |
| 2010 | 219614 | 652455 | 54638 | 31589 | 2979 | 1651 | 618 | 150 | 224 |
| 2011 | 57933 | 81604 | 255899 | 17089 | 10585 | 1071 | 611 | 236 | 146 |
| 2012 | 41955 | 21527 | 33558 | 90543 | 6523 | 4333 | 452 | 266 | 169 |
| 2013 | 573690 | 15590 | 8501 | 10686 | 30922 | 2389 | 1635 | 176 | 172 |
| 2014 | 130438 | 213170 | 6340 | 2922 | 3956 | 12278 | 978 | 690 | 150 |
| 2015 | 387698 | 48468 | 86356 | 2157 | 1070 | 1554 | 4970 | 408 | 355 |
| 2016 | 75598 | 144060 | 19963 | 30681 | 827 | 440 | 658 | 2170 | 339 |
| 2017 | 79577 | 28091 | 56762 | 6319 | 10412 | 301 | 165 | 254 | 982 |
| 2018 | 873064 | 29569 | 10955 | 17494 | 2085 | 3684 | 110 | 62 | 477 |
| 2019 | 253861 | 324410 | 10935 | 2940 | 4987 | 637 | 1161 | 36 | 180 |

Table 7.4.8. Haddock in 7.bc-ek. Stock Summary: weights in tonnes; CatchPred is predicted catch from ASAP; recruitment at age zero ( ${ }^{\prime} 1000$ ); $F_{\text {bar }}$ ages 3-5.

| Year | Lan | Dis | Cat | CatPred | Tsb | Ssb | SsbCv | Recr | RecrCv | Fbar | FbarCv |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 3348 | 1208 | 4556 | 4675 | 16818 | 7550 | 0.210 | 111802 | 0.211 | 1.087 | 0.246 |
| 1994 | 4131 | 1886 | 6017 | 5398 | 28097 | 8114 | 0.219 | 383799 | 0.182 | 1.064 | 0.236 |
| 1995 | 4470 | 2218 | 6688 | 6455 | 46077 | 7509 | 0.193 | 533832 | 0.160 | 0.838 | 0.259 |
| 1996 | 6756 | 4309 | 11065 | 12205 | 46746 | 19979 | 0.180 | 150687 | 0.196 | 0.830 | 0.258 |
| 1997 | 10827 | 2883 | 13710 | 13620 | 37906 | 29116 | 0.156 | 76803 | 0.221 | 0.703 | 0.256 |
| 1998 | 7928 | 934 | 8862 | 9988 | 31453 | 22775 | 0.157 | 160517 | 0.192 | 0.746 | 0.245 |
| 1999 | 4970 | 586 | 5556 | 6042 | 32454 | 14458 | 0.157 | 423278 | 0.181 | 0.513 | 0.290 |
| 2000 | 7499 | 2503 | 10002 | 10870 | 45714 | 17833 | 0.165 | 397699 | 0.204 | 0.655 | 0.279 |
| 2001 | 9278 | 3418 | 12696 | 16561 | 54818 | 29759 | 0.165 | 453878 | 0.174 | 0.696 | 0.304 |
| 2002 | 6488 | 7073 | 13561 | 23529 | 72031 | 36655 | 0.201 | 801131 | 0.138 | 1.238 | 0.236 |
| 2003 | 8292 | 9456 | 17748 | 17458 | 66415 | 25954 | 0.164 | 219000 | 0.154 | 0.653 | 0.268 |
| 2004 | 8777 | 6750 | 15527 | 21918 | 63781 | 43868 | 0.139 | 284062 | 0.129 | 0.774 | 0.246 |
| 2005 | 6787 | 5191 | 11978 | 15012 | 55040 | 30228 | 0.150 | 274063 | 0.123 | 0.814 | 0.232 |
| 2006 | 5593 | 2484 | 8077 | 10523 | 47753 | 23989 | 0.135 | 202132 | 0.138 | 0.508 | 0.287 |
| 2007 | 6781 | 2739 | 9520 | 8670 | 65604 | 25527 | 0.127 | 711866 | 0.105 | 0.406 | 0.268 |
| 2008 | 7455 | 11187 | 18642 | 15521 | 77948 | 24448 | 0.127 | 370642 | 0.129 | 0.724 | 0.177 |
| 2009 | 9608 | 9080 | 18688 | 16111 | 137393 | 36875 | 0.106 | 1752675 | 0.088 | 0.570 | 0.186 |
| 2010 | 10262 | 16547 | 26809 | 25630 | 130467 | 38136 | 0.116 | 218799 | 0.136 | 0.591 | 0.182 |
| 2011 | 12879 | 14378 | 27257 | 27938 | 99511 | 86470 | 0.089 | 57476 | 0.189 | 0.463 | 0.190 |
| 2012 | 18376 | 10191 | 28567 | 25780 | 74114 | 69258 | 0.097 | 41446 | 0.201 | 0.572 | 0.160 |
| 2013 | 13424 | 2085 | 15509 | 13459 | 75657 | 43200 | 0.108 | 562000 | 0.099 | 0.491 | 0.180 |
| 2014 | 9854 | 3177 | 13031 | 12104 | 65868 | 28258 | 0.133 | 125695 | 0.156 | 0.509 | 0.196 |
| 2015 | 8545 | 6694 | 15239 | 14627 | 76360 | 43106 | 0.104 | 399920 | 0.114 | 0.478 | 0.200 |
| 2016 | 7594 | 10337 | 17931 | 16555 | 66624 | 33794 | 0.119 | 75314 | 0.198 | 0.619 | 0.183 |
| 2017 | 8097 | 7975 | 16072 | 15819 | 49119 | 38018 | 0.113 | 78760 | 0.223 | 0.627 | 0.201 |
| 2018 | 7109 | 5436 | 12546 | 12209 | 74659 | 24239 | 0.152 | 869103 | 0.216 | 0.773 | 0.286 |
| 2019* | NA | NA | NA | NA | NA | 14205 | NA | 252713 | NA | 0.773 | NA |

* GM recruitment and mean $F$ last over the three years.

Table 7.4.9. Haddock in 7.bc-ek. Input values for short-term forecast. Note that Sel and CWt refer to the landings and DSel and DCWt refer to the discards. Numbers in thousands; Weights in kg.

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt | DSel | DCWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 252713 | 0.99 | 0 | 0 | 0 | 0.058 | 0.000 | 0.000 | 0.000 | 0.061 |
| 1 | 322939 | 0.72 | 0 | 0 | 0 | 0.217 | 0.013 | 0.453 | 0.231 | 0.224 |
| 2 | 10758 | 0.60 | 1 | 0 | 0 | 0.368 | 0.202 | 0.688 | 0.433 | 0.335 |
| 3 | 2872 | 0.50 | 1 | 0 | 0 | 0.660 | 0.483 | 0.899 | 0.190 | 0.527 |
| 4 | 4916 | 0.43 | 1 | 0 | 0 | 1.031 | 0.608 | 1.212 | 0.065 | 0.770 |
| 5 | 569 | 0.40 | 1 | 0 | 0 | 1.494 | 0.657 | 1.465 | 0.016 | 1.118 |
| 6 | 1038 | 0.37 | 1 | 0 | 0 | 1.910 | 0.666 | 1.816 | 0.007 | 1.534 |
| 7 | 32 | 0.36 | 1 | 0 | 0 | 2.101 | 0.657 | 1.905 | 0.016 | 1.658 |
| 8 | 164 | 0.34 | 1 | 0 | 0 | 2.311 | 0.669 | 2.007 | 0.004 | 2.817 |
| 2020 |  |  |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt | DSel | DCWt |
| 0 | 252713 | 0.99 | 0 | 0 | 0 | 0.058 | 0.000 | 0.000 | 0.000 | 0.061 |
| 1 | 93902 | 0.72 | 0 | 0 | 0 | 0.217 | 0.013 | 0.453 | 0.231 | 0.224 |
| 2 | 118710 | 0.60 | 1 | 0 | 0 | 0.368 | 0.202 | 0.688 | 0.433 | 0.335 |
| 3 | 2848 | 0.50 | 1 | 0 | 0 | 0.660 | 0.483 | 0.899 | 0.190 | 0.527 |
| 4 | 804 | 0.43 | 1 | 0 | 0 | 1.031 | 0.608 | 1.212 | 0.065 | 0.770 |
| 5 | 1476 | 0.40 | 1 | 0 | 0 | 1.494 | 0.657 | 1.465 | 0.016 | 1.118 |
| 6 | 176 | 0.37 | 1 | 0 | 0 | 1.910 | 0.666 | 1.816 | 0.007 | 1.534 |
| 7 | 331 | 0.36 | 1 | 0 | 0 | 2.101 | 0.657 | 1.905 | 0.016 | 1.658 |
| 8 | 64 | 0.34 | 1 | 0 | 0 | 2.311 | 0.669 | 2.007 | 0.004 | 2.817 |
| 2021 |  |  |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt | DSel | DCWt |
| 0 | 252713 | 0.99 | 0 | 0 | 0 | 0.058 | 0.000 | 0.000 | 0.000 | 0.061 |
| 1 | 93902 | 0.72 | 0 | 0 | 0 | 0.217 | 0.013 | 0.453 | 0.231 | 0.224 |
| 2 | 34518 | 0.60 | 1 | 0 | 0 | 0.368 | 0.202 | 0.688 | 0.433 | 0.335 |
| 3 | 31432 | 0.50 | 1 | 0 | 0 | 0.660 | 0.483 | 0.899 | 0.190 | 0.527 |
| 4 | 798 | 0.43 | 1 | 0 | 0 | 1.031 | 0.608 | 1.212 | 0.065 | 0.770 |
| 5 | 242 | 0.40 | 1 | 0 | 0 | 1.494 | 0.657 | 1.465 | 0.016 | 1.118 |
| 6 | 457 | 0.37 | 1 | 0 | 0 | 1.910 | 0.666 | 1.816 | 0.007 | 1.534 |
| 7 | 56 | 0.36 | 1 | 0 | 0 | 2.101 | 0.657 | 1.905 | 0.016 | 1.658 |
| 8 | 128 | 0.34 | 1 | 0 | 0 | 2.311 | 0.669 | 2.007 | 0.004 | 2.817 |

Table 7.4.10. Haddock in 7.bc-ek. Single-option output of the short-term forecast ( $F=$ mean F2013-2015). Numbers in thousands, weights in tonnes.

2019

| Age | F | CatchNos | Yield | DF | DCatchNos | DYield | StockNos | Biomass | SSNos | SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.000 | 0 | 0 | 0.000 | 0 | 0 | 252713 | 14657 | 0 | 0 |
| 1 | 0.015 | 3041 | 1376 | 0.266 | 54257 | 12136 | 322939 | 70185 | 0 | 0 |
| 2 | 0.232 | 1379 | 948 | 0.497 | 2959 | 990 | 10758 | 3962 | 10758 | 3962 |
| 3 | 0.555 | 902 | 810 | 0.218 | 354 | 187 | 2872 | 1896 | 2872 | 1896 |
| 4 | 0.698 | 1996 | 2419 | 0.075 | 213 | 164 | 4916 | 5068 | 4916 | 5068 |
| 5 | 0.754 | 253 | 371 | 0.018 | 6 | 7 | 569 | 851 | 569 | 851 |
| 6 | 0.765 | 473 | 859 | 0.008 | 5 | 8 | 1038 | 1982 | 1038 | 1982 |
| 7 | 0.754 | 14 | 28 | 0.019 | 0 | 1 | 32 | 67 | 32 | 67 |
| 8 | 0.768 | 76 | 153 | 0.005 | 0 | 1 | 164 | 379 | 164 | 379 |
| Total | 1.000 | 8134 | 6964 | 0.000 | 57794 | 13494 | 596001 | 99047 | 20349 | 14205 |
| 2020 |  |  |  |  |  |  |  |  |  |  |
| Age | F | CatchNos | Yield | DF | DCatchNos | DYield | StockNos | Biomass | SSNos | SSB |
| 0 | 0.000 | 0 | 0 | 0.000 | 0 | 0 | 252713 | 14657 | 0 | 0 |
| 1 | 0.015 | 884 | 400 | 0.266 | 15777 | 3529 | 93902 | 20408 | 0 | 0 |
| 2 | 0.232 | 15214 | 10462 | 0.497 | 32657 | 10929 | 118710 | 43725 | 118710 | 43725 |
| 3 | 0.555 | 894 | 804 | 0.218 | 351 | 185 | 2848 | 1880 | 2848 | 1880 |
| 4 | 0.698 | 327 | 396 | 0.075 | 35 | 27 | 804 | 829 | 804 | 829 |
| 5 | 0.754 | 656 | 961 | 0.018 | 16 | 18 | 1476 | 2206 | 1476 | 2206 |
| 6 | 0.765 | 80 | 146 | 0.008 | 1 | 1 | 176 | 337 | 176 | 337 |
| 7 | 0.754 | 149 | 285 | 0.019 | 4 | 6 | 331 | 695 | 331 | 695 |
| 8 | 0.768 | 30 | 60 | 0.005 | 0 | 0 | 64 | 148 | 64 | 148 |
| Total | 1.000 | 18234 | 13514 | 0.000 | 48841 | 14695 | 471024 | 84885 | 124409 | 49820 |
| 2021 |  |  |  |  |  |  |  |  |  |  |
| Age | F | CatchNos | Yield | DF | DCatchNos | DYield | StockNos | Biomass | SSNos | SSB |
| 0 | 0.000 | 0 | 0 | 0.000 | 0 | 0 | 252713 | 14657 | 0 | 0 |
| 1 | 0.015 | 884 | 400 | 0.266 | 15777 | 3529 | 93902 | 20408 | 0 | 0 |
| 2 | 0.232 | 4424 | 3042 | 0.497 | 9496 | 3178 | 34518 | 12714 | 34518 | 12714 |
| 3 | 0.555 | 9868 | 8868 | 0.218 | 3872 | 2042 | 31432 | 20745 | 31432 | 20745 |
| 4 | 0.698 | 324 | 393 | 0.075 | 35 | 27 | 798 | 822 | 798 | 822 |
| 5 | 0.754 | 107 | 157 | 0.018 | 3 | 3 | 242 | 361 | 242 | 361 |
| 6 | 0.765 | 208 | 378 | 0.008 | 2 | 3 | 457 | 873 | 457 | 873 |
| 7 | 0.754 | 25 | 48 | 0.019 | 1 | 1 | 56 | 118 | 56 | 118 |
| 8 | 0.768 | 59 | 119 | 0.005 | 0 | 1 | 128 | 295 | 128 | 295 |
| Total | 1.000 | 15899 | 13405 | 0.000 | 29186 | 8784 | 414246 | 70993 | 67631 | 35928 |

Table 7.4.11. Haddock in 7.bc-ek. Management options table. Weights in tonnes.
$\left.\begin{array}{lcccccccccc}\hline \text { Fmult } & \text { Catch20 } & \text { Land20 } & \text { Dis20 } & \text { FCatch20 } & \text { FLand20 } & \text { FDis20 } & \text { SSB21 } & \begin{array}{l}\text { Change } \\ \text { SSB2019 }\end{array} & \begin{array}{l}\text { Change } \\ \text { TAC } \\ \text { 2019 }\end{array} & \begin{array}{l}\text { Change } \\ \text { TAC } \\ \text { 2019 to } \\ \text { Land }\end{array} \\ \begin{array}{l}\text { Change- } \\ \text { Advice } \\ \text { 2019 to } \\ \text { 2020 }\end{array} \\ \text { 2020 } \\ \text { Catch }\end{array}\right)$


Figure 7.4.1. International haddock landings by ICES rectangle (all gears; 2016; data from https://stecf.jrc.ec.eu-ropa.eu/data-dissemination).


Figure 7.4.2. a. Haddock in 7.bc-ek. Official Ices landings and TAC of haddock in 7.b-k.


Figure 7.4.2 b. Haddock in 7.bc-ek. Recent working group landings and quota by country.


Figure 7.4.3. Haddock in 7.bc-ek. Effort ('1000h) of the Irish Otter trawl fleets, the French demersal otter trawl fleet and for UK trawl fleet (effort in fishing days, rescaled to other fleets) and lpue (kg/h) for the Irish and French fleets.


Figure 7.4.4 Proportional representation of landings relative to catch (discards + landings) by age, 2013-2018.


Figure 7.4.5. Haddock in 7.bc-ek. Discarding by number by age class 2000-2018 (grey = landings, white $=$ discards).


Figure 7.4.6. Haddock in 7.bc-ek. Proportion of discards by age (left) and year (right).


Figure 7.4.7. Haddock in 7bc-ek. Distribution sampled and unsampled the catches by country and gear (left) and by age (right). Note that both France and Ireland allocated age data to most unsampled strata before uploading to InterCatch.


Figure 7.4.8. Haddock in 7.bc-ek. Raw stock weights-at-age (left) and the three-year running average stock weights (right).


Figure 7.4.9. Haddock in 7.bc-ek. Log standardised indices of tuning fleets by year. The FRA-IRL-IBTS survey is the combined French EVHOE Q4 WIBTS and Irish IGFS Q4 WIBTS survey. The IRL-GAD commercial tuning fleet is the Irish gadoid fleet in 7.gj.


Figure 7.4.10. Haddock in 7.bc-ek. Log standardised indices of tuning fleets by cohort.


Figure 7.4.11. Haddock in 7.bc-ek. Scatterplot matrix of log indices of cohorts at different ages.


Figure 7.4.12. Haddock in 7.bc-ek. Residuals of the proportions-at-age index observed to predicted, (upper) and standardised relative to predicted (lower).


Figure 7.4.13. Haddock in 7.bc-ek. Observed and predicted catches.


Figure 7.4.14. Haddock in 7.bc-ek. Index proportions-at-age residuals (observed and predicted).


Figure 7.4.15. Haddock in 7.bc-ek. Observed and predicted index cpue.


Fbar 3-5


Catch


Fbar 3-5


SSB


Recruits age 0


SSB


Recruits age 0


Figure 7.4.16. Haddock in 7.bc-ek. ASAP assessment stock summary plots.

- Top four plots: Lines represent the ASAP assessment with standard deviation shaded. Forecast/ assumed values are given by open circles.
- Bottom four plots: Comparison of ASAP (black) and XSA (red) assessments. For ASAP the $\mathrm{F}_{\text {bar }}$ range was 3-5 while for XSA 2-5. The natural mortality assumption for the ASAP is much higher for young ages than the assumed $M$ for the historic XSAs, resulting in a higher estimate of recruitment.

Landings yield 2020
SSB 2021


Figure 7.4.17. Haddock 7bc-ek. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.


Figure 7.4.18. Haddock in 7.bc-ek. Retrospective analysis of the final ASAP run. Note that the survey index only started in 2003.


Figure 7.4.19. Haddock 7bc-ek. Retrospective analysis of the stock assessment, comprising the ASAP assessment of the stock status up to the most recent catch year (2018) and current year stock status made on the intermediate year assumptions (Section 12.13) . Note that the survey index only started in 2003.


Figure 7.4.20. Haddock 7bc-ek. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.

## 10 Megrim (Lepidorhombus ssp.) in divisions 4.a and 6.a (northern North Sea, West of Scotland)

## Type of assessment in 2018

Update of 2018 assessment with new landings and survey data. The model used to carry out the assessment is the Schaefar Surplus production process model in R and Winbugs.

ICES advice applicable to 2019 and 2020
ICES advises that when the EU multiannual plan (MAP) is applied, catches in 2019 that correspond to the $F$ ranges in the plan are between 6450 tonnes and 8350 tonnes.

### 10.1 General

## Stock description and management units

Megrim stock structure is uncertain and historically the Working Group has considered megrim populations in $6 . a$ and $6 . \mathrm{b}$ as separate stocks. The review group questioned the basis for this in 2004. Data collected during an EC study contract (98/096) on the 'Distribution and biology of anglerfish and megrim in the waters to the West of Scotland' showed significantly different growth parameters and significant population structure difference between megrim sampled in 6.a and 6.b (Anon, 2001). Spawning fish occur in both areas but whether these populations are reproductively isolated is not clear. As noted by WGNSDS (2008), megrim in 4 .a has historically not been considered by ICES and WGNSDS (2008). Since 2009, data from 4 and 2.a are included in this report, but international catch and weight-at-age data for 4, prior 2006 were not available to the working group or WKFLAT (2011). Given that there is little evidence to suggest that megrim in $6 . a$ and 4.a are separate stocks, based on a visual inspection of the spatial distribution of commercial landings and fishery-independent survey data, WKFLAT (2011) concluded that megrim in $6 . a$ and 4 .a should be considered as a single stock. This has subsequently been supported through recent genetic studies (MacDonald and Prieto, 2012) indicating that there is one stock consisting of divisions $4 . a$ (northern North Sea) and 6.a (West of Scotland) and another separate stock in Division 6.b (Rockall).


| Species: | Megrims <br> Lepidarhombus spp. |  | Zone: | Union waters of $2 a$ and 4 (LEZ/2AC4-C) |
| :---: | :---: | :---: | :---: | :---: |
| Belgium |  | 8 |  |  |
| Denmark |  | 7 |  |  |
| Germany |  | 7 |  |  |
| France |  | 41 |  |  |
| The Netherlands |  | 33 |  |  |
| United Kingdom |  | 2430 |  |  |
| Union |  | 2526 |  |  |
| TAC |  | 2526 |  | Analytical TAC <br> Article 7(2) of this Regulation applies |
| Species: | Megrims <br> Lepidarhombus spp. |  | Zone: | Union and international waters of 5 b; 6; international waters of 12 and 14 (LEZ/56-14) |
| Spain |  | 617 |  |  |
| France |  | 2407 |  |  |
| Ireland |  | 704 |  |  |
| United Kingdom |  | 1704 |  |  |
| Union |  | 5432 |  |  |
| TAC |  | 5432 |  | Analytical TAC <br> Article 7(2) of this Regulation applies |

2018 TAC for 6, EC waters of 5.b and International waters of 12 and 14 (lower) and TAC for 4 and 2.a (upper).

| Species: | Megrims <br> Lepidorhombus spp. |  | Zone: | Union waters of $2 a$ and 4 (LEZ/2AC4- |
| :---: | :---: | :---: | :---: | :---: |
| Belgium |  | 9 |  |  |
| Denmark |  | 7 |  |  |
| Germany |  | 7 |  |  |
| France |  | 47 |  |  |
| The Netherlands |  | 37 |  |  |
| United Kingdom |  | 2780 |  |  |
| Union |  | 2887 |  |  |
| TAC |  | 2887 |  | Analytical TAC <br> Article 7(2) of this Regulation applies |
| Species: | Megrims <br> Lepidorhombus spp. |  | Zone: | Union and international waters of 5 b; 6 ; international waters of 12 and 14 $(\mathrm{LEZ} / 56-14)$ |
| Spain |  | 657 |  |  |
| France |  | 2563 ( $\left.{ }^{( }\right)$ |  |  |
| Ireland |  | 749 |  |  |
| United Kingdom |  | $1813{ }^{(2)}$ |  |  |
| Union |  | 5782 |  |  |
| TAC |  | 5782 |  | Analytical TAC <br> Article 7(2) of this Regulation applies |

2019 TAC for 6 , EC waters of 5. b and International waters of 12 and 14 (lower) and TAC for 4 and 2.a (upper).

The uptake of the 2018 TAC for ICES Division 6 and EU waters of $5 . b$ was $33 \%$. Uptake varied considerably between countries. The UK, which holds much of the quota allocation, utilised only $23 \%$ of its allocation.

In ICES areas 4 and 2.a, the TAC was overshot by $13 \%$ in 2018. The majority of available TAC ( $96 \%$ ) is allocated to the UK, who take $87 \%$ of it.

## Fisheryin 2018

## Landings

Official landings data for each country together with Working Group best estimates of landings from 6.a are shown in Table 13.2 and for 4.a in Table 13.3. To estimate ICES landings we take InterCatch estimates and, if unavailable, we use official estimates. There are a few discrepancies with the estimates, for example there are no Danish data in InterCatch for 2017 and 2018. There are often minor differences between official data and InterCatch for most countries.

Catches of megrim comprise two species, Lepidorhombus whiffiagonis and L. boscii. Information available to the Working Group indicates that L. boscii, are a negligible proportion of the Scottish and Irish megrim catch (Kunzlik et al.,1995; Anon, 2001). Commercial catches are dominated by female megrim, typically $90 \%$ of the total catch.

The InterCatch catch estimate is 3258 tonnes, and the ICES catch estimate for $6 . a$ and $4 . a$. is 3003 tonnes. The total ICES landings are well below the total TAC covering the fished areas of 4.a-6.a.

## Discards

Raised discard data were made availableby Scotland, France, Spain and Ireland (6.a and 4.a) and Scotland and Ireland (6.a). Scottish data give a discard rate of 7.9\%, Irish discards were 10.2\%, Spanish discards were $0.3 \%$, and French discards were $13.9 \%$ by weight. Total discards were estimated to be 255 t or $7.8 \%$ by weight for the stock area in 2018. We assume no discards for Denmark, Netherlands, and Germany.

A linear decline in discards from 30 to $15 \%$ over time between 1985 and 2012 is assumed in the stock assessment. From 2013 onwards discard data have taken from InterCatch, there is no deviation from the agreed stock annex (see Annex 2).

## Catch

A breakdown of 2018 catch by main gear type in InterCatch is given below:

| Catch | Landings |  | Discards |  |  |  |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| 3258 tonnes | Finfish trawls | Nephrops trawls | Other Gears | Finfish trawls | Nephrops trawls | Other Gears |
|  | $96.6 \%$ | $0.52 \%$ | $2.9 \%$ | $96 . .5 \%$ | $3.4 \%$ | $0.029 \%$ |
|  |  | 3003 tonnes | 255 tonnes |  |  |  |

## Surveys

Indices from six fishery-independent surveys are used in the assessment. The surveys are outlined in Table 13.1 below and details can be viewed in the stock annex.

Table 13.1. Summary indices used for surplus production model.

| NUMBER | SURVEY | NATIONALITY | AREA | TIME-SERIES | DEPTH RANGE(M) |
| :--- | :--- | :--- | :--- | :---: | :---: |
| 1 | Sco-IBTS-Q3 | Scotland | $4 . a$ | $1987-2017$ | $50-250$ |
| 2 | Sco-IBTS-Q1 | Scotland | $4 . a$ | $1987-2017$ | $50-250$ |
| 3 | ScoGFS-WIBTS-Q1 | Scotland | Scotland | $6 . a$ | $1986-2010$ |
| 4 | ScoGFS-WIBTS-Q4 | Scotland | $6 . a / 4 . a$ | $2005-2017$ | $40-400$ |
| 5 | SAMISS-Q2 | Ireland | $6 . a$ | $2005-2017$ | $50-1050$ |
| 6 | IAMISS-Q2 |  |  | $50-850$ |  |

The SAMISS and IAMISS surveys were combined for assessment purposes.
Figures 13.1 to 13.5 present the megrim biomass maps for the AMISS and IBTS surveys. The AMISS bubble plots show and increasing abundance over time throughout the area over the time-series. The abundance in 6.a was particularly high in 2013 and a similar high abundance occurred in $4 . \mathrm{a}$ in 2014 (Figure 13.1). Figures 13.2. (Sco-IBTS-Q3 4.a) and 13.3 (Sco-IBTS-Q1 4.a) show the large increase in biomass over time in the northern North Sea. Biomass in the southern North Sea remains quite low.

Figures 13.4 (Sco-GFS-Q1 4.a) and 13.5 (Sco-GFS-Q4 4.a) also show an increase in biomass over the time-series. However, the survey design and ground gear changed after 2010 so this should be taken into account when interpreting the plots. Data from these data were truncated from the time-series going into the assessment.

### 10.2 Estimation of survey cpue indices

## C pue trends of survey data

The data from the IBTS surveys exhibit a relatively large proportion of zeros, therefore the delta method of Stefánsson (1996) was used to generate indices. This method (delta-gamma model) comprises fitting two generalized linear models. The first model (binomial GLM) is used to obtain the proportion of non-zero tows, and is fit to the data coded as 1 or 0 , if the tow contained a positive or zero cpue, respectively. The second model is fit to the positive only cpue data using a gamma or lognormal GLM.

At WGCSE 2017, it was discovered that previous delta-gamma cpue estimations had included the full time-series for the $6 . a$ surveys, when fitting the model to those surveys. In 2019, these again generate a slightly different cpue index Figure 13.6. The truncated series was used in the 2019 assessment since fitting to the full series would be inappropriate.

The biomass trend for the AMISS survey is shown in Figure 13.7. There is a weakly increasing trend over time with year effects evident in $6 . a$ in 2013 and 4.a in 2014, while a notable increase in 2016 and 2017 was followed by a decrease in 2018. The biomass trends for the four IBTS surveys are shown in Figure 13.8.

## Commercial cpue

Commercial cpue data have not been updated compared to last year and are not used in the assessment.

### 10.3 Stock assessment

The input data for the stock assessment are given in Table 13.4 this comprises of a time-series from all six surveys and ICES catch estimates for this stock.

## 2019 Final run

The Pearson residuals diagnostic plots for the final assessment are shown in Figure 13.9. The residuals for the two 6. a surveys and the AMISS survey are fairly randomly dispersed around zero. A trend in the residuals is evident for the two 4 .a surveys is evident with increasing positive residuals in the last decade.

The prior and posterior distributions for the parameters in the final model fit, are shown in Figure 13.10. The priors are given in Table 13.5. The posterior distributions are similar to previous years' assessments. The posterior parameter estimates for the final assessment model are given in Table 13.6. These are similar to recent assessments.

Figure 13.11 shows the final model fits to the cpue series and the estimates of total biomass and harvest ratio. The fits to the $6 . a$ and AMISS surveys are reasonable. The fits to the $4 . a$ surveys show that the model is not fitting well to those surveys in recent years. This issues needs to be examined further in the next benchmark.

Figure 13.12 compares the assessment results of the model fitted with to a cpue generated using the full time-series of the 6 .a surveys and a model with the truncated cpue series. This indicates that the impact of fitting the model to the full time-series of delta-gamma cpues for 6 .a instead of the truncated time-series was minimal, mainly effecting the early part of the time-series.

The time-series of $B / B_{\text {MSY }}$ and $F / F_{\text {MSY }}$ landings and discards used in the final assessment are given in Table 13.7.

## Comparison with previous assessments

Figure 13.13 compares the final assessment with those conducted by WGCSE at previous meetings. The 2019 assessment revised down recent biomass estimates and up recent fishing mortality estimates. There is also some deviations in the historic estimates of F and Biomass around 2000. This is linked to the use of the truncated 6.a surveys to derive the delta-gamma cpues to input to the assessment model.

To evaluate evidence of possible bias in the assessment population metrics, a Mohn's Rho analysis resulted in values of -0.092 for Fbar and 0.088 for biomass. ICES considers a valuegreater than 0.20 to be unacceptably high.

## State of the stock

The state of the stock has not changed since last year. Fishing mortality has been below Fmsy for almost the full time-series and has an overall declining trend since the late 1990s. Biomass has consistently been above MSY $B_{\text {trigger }}$ and shows an increasing trend since 2005. The stock in 2018 is estimated to be 1.7 times $B_{\text {MSY }}$. The fishing mortality in 2018 is estimated to be have been $35 \%$ of Fmsy.

### 10.4 Short-term projections

Short-term projections have been updated according to the method set out in the stock annex. The basis for the catch options is given in Table 13.8.

The management option table is given in Table 13.9. Fishing at $\mathrm{F}_{\text {MSY }}$ in 2019 is projected to result in total catches of 8350 t (landing of 7684 t and discards of 666 t ) and a Biomass of 1.4 times BMSY in 2021.

### 10.5 Biological reference points

## Precautionary ap proach reference points

FmSY, BMSY and the yield at MSY are all directly estimated in the model. It should be noted that these will vary when new survey and catch information is added. $\mathrm{B}_{\mathrm{pa}}$ and Blim are defined as $50 \% \mathrm{~B}_{\text {MSY }}$ and $30 \% \mathrm{~B}_{\text {MSY }}$ respectively. $\mathrm{F}_{\text {lim }}$ is defined as $1.7 \mathrm{~F}_{\text {MSY }}$ and is the F that drives the stock to $\mathrm{B}_{\mathrm{lim}}$ assuming $\mathrm{Blim}_{\mathrm{lim}}=30 \% \mathrm{Bmsy}$. The derivation is given below:

```
P=rB(l-B/K)
The surplus productivity associated with Blim is:
P
The corresponding F is:
F}\mp@subsup{F}{lim}{}=r\mp@subsup{B}{lim}{lm}(1-\mp@subsup{B}{\mathrm{ lim }}{}/K)/\mp@subsup{B}{lim}{}=r(1-\mp@subsup{B}{lim}{l}/K
B _ { l i m } = 0 . 3 B _ { M S Y } = 0 . 3 \mathrm { K } / 2
Flim}=r(1-0.3K/(2K))=r(1-0.3/2)=0.85r
F _ { M S Y } = r / 2 , \text { let x denote the proportionality between } F _ { M S Y } \text { and } F _ { \text { lim} }
x FMSY }=\mp@subsup{F}{\mathrm{ lim}}{
x(r/2)=0.85r
x=2*0.85
x=1.7
```


## MSY reference points

In 2015, ICES provided precautionary FMSY ranges that are derived to deliver no more than a $5 \%$ reduction in long-term yield compared with MSY. Details of this analysis are given in WKMSYREF3 (ICES, 2015) and the derivations are given below.

|  | MSY Flower ${ }^{\text {b }}$ | $\mathrm{F}_{\text {MSY }}{ }^{\text {b }}$ | MSY upper $^{\text {b }}$ ( with AR | MSY ${ }^{\text {trigger }}$ |
| :---: | :---: | :---: | :---: | :---: |
| Megrim in divisions 4.a and 6.a | $0.39 \times r^{\text {d }}$ ) | $\mathrm{r} / 2 \mathrm{~d})$ | $\mathrm{r} / 2 \mathrm{~d})$ | K d) |

Because the stock has been fished below FMSY for more than ten years the WG considered it appropriate to set the MSY $B_{\text {trigger }}=B_{\text {MSY }}$ according to the ICES guidelines (ICES, 2017).

## Uncertainties and bias in assessment and forecast

The model estimates of B and F have large uncertainty. Despite this, there is a low probability that SSB is below MSY $\mathrm{B}_{\text {Trigger }}$ and a high probability that F is below $\mathrm{F}_{\text {MSY }}$.

The reference points are re-estimated within the assessment. The change between 2017 and 2018 reference points is less than hasbeen previously seen, and results in a notable rescaling of relative stock status. However, in absolute terms, stock trends are consistent with those of previous years.

The biomass time-series from surveys has increasing uncertainty boundaries as the index increases. This results in uncertainty bounds in the model estimates show a contraction from the 2018 assessment.

Owing to incomplete discard data, historical discard rates (1985-2012) are assumed to have declined, from $30 \%$ at the beginning of the time-series, to an estimate of $15 \%$ in 2012. The evaluation of current stock status is robust to this assumption. Estimates since 2013 are based on observed discards.

## Recommendation for next benchmark

This stock was subject to an inter-benchmark in 2012 (IBP-MEG, 2012). Due to incomplete age data, particularly for $4 . a$, a Bayesian state-space surplus production model was chosen as the final assessment model. Subsequent update assessments have highlighted a problem fitting to the 4 .a surveys which needs to be examined in a future benchmark.

WGCSE recommends the following explorations:

- The AMISS survey should be merged into one continuous index. The length data for the index should also be examined.
- TheSco6.a Q1/Q4 WIBTS 2011+:theSco6.a Q1/Q4 WIBTS survey time-series should also be examined for re-introduced into the assessment as a new time-series. There may also be scope to integrate the IGFS.
- Available length and age-structured data should be compiled for this stock.
- Length or age-structured assessment models could be explored.

Once sufficient progress has been made on the points above, WGCSE will suggest a benchmark schedule.

## Management considerations

Megrim is a bycatch species in the mixed demersal trawl in divisions 6.a and 4.a. Management measures for other species have constrained the fishery and reduced effort and fishing mortality on megrim. The general increase in mesh size in 6 and 4 since 2010 has also benefited the stock.

The TAC in 6 has not been fully utilised. However, the uptake rate is country specific, with some Member States reporting landings above their quota in the North Sea. Partial quota uptake by individual Member States may be linked to reduction in effort rather than reflective of a reduction in biomass. The TAC and assessment area are incompatible. There are two separate TAC areas covering ICES areas 6 and 4, whereas the assessment covers ICES divisions 6.a and 4.a combined. Due consideration of the inconsistency between management and assessment area is required when setting fishing opportunities for this stock and the separate 6.b Rockall stock. ICES (2013) have advised the EC that the TAC areas should be consistent with the assessment area and that ICES has no basis on how to split the catch advice so that it is consistent with the TAC areas.

### 10.6 References

ICES. 2015. Report of the Joint ICES-MYFISH Workshop to consider the basis for Fmsy ranges for all stocks (WKMSYREF3), 17-21 November 2014, Charlottenlund, Denmark. ICES CM 2014/ACOM:64. 156 pp.
ICES. 2012. Report of the Inter-benchmark Protocol for Megrim in Subarea IV and Division IVa (IBPMeg), 2-6 April 2012. By correspondence. ICES CM 2012/ACOM:67. 23 pp.

ICES. 2011. Report of the Benchmark Workshop on Flatfish (WKFLAT), 1-8 February 2011, Copenhagen, Denmark. ICES CM 2011/ACOM:39. 257 pp.
ICES. 2011. Report of the Working Group for Celtic Seas Ecoregion (WGCSE), 11-19 May 2011, Copenhagen, Denmark. ICES CM 2011/ACOM:12.1573 pp.

Kunzlik, P. A., A. W. Newton and A. W. Jermyn. 1995. Exploitation of monks (Lophius spp.) and megrims (Lepidorhombus spp.) by Scottish fishermen in ICES Division VIa (West of Scotland). Final report EU FAR contract MA-2-520.
Laurenson, C. and MacDonald, P. 2008. Collection of fisheries and biological data on megrim in ICES Subarea IVa. Scottish Industry Science Partnership Report No 05/08.

Meyer and Millar. 1999. BUGS in Bayesian stock assessments. Canadian Journal of Fisheries and Aquatic Sciences; Jun 1999; 56, 6; Canadian Periodicals. pp. 1078.
Reid, D.G., Allen, V.J., Bova, D.J., Jones, E.G., Kynoch, R.J., Peach, K.J., Fernandes, P.G. and Turrell, W.R. 2007. Angler fish catchability for swept area abundance estimates in a new survey trawl. ICES J. Mar. Sci. 64 .

Stefánsson, G. 1996. Analysis of groundfish survey abundance data: combining the GLM and delta approaches. ICES Journal of Marine Science, 53, 577-588.

Table 13.2. Megrim in Subarea 6.a. Nominal catch ( $t$ ) of Megrim West of Scotland, as officially reported to ICES and WG best estimates of landings. The shaded cells show updates in official data compared with last year.

|  |  | $\begin{aligned} & \text { 쁜 } \\ & \text { 든 } \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \text { 든 } \\ & \underline{\underline{0}} \end{aligned}$ |  | $\begin{aligned} & \text { 드추 } \\ & \text { ì } \end{aligned}$ |  |  | $\underset{ }{〕}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 0 | 398 | 317 | 0 | 91 | 25 | 1093 | - | 1924 | 2210 |
| 1991 | 1 | 455 | 260 | 0 | 48 | 167 | 1223 | - | 2154 | 2432 |
| 1992 | 0 | 504 | 317 | 0 | 25 | 392 | 887 | - | 2125 | 2549 |
| 1993 | 0 | 517 | 329 | 0 | 7 | 298 | 896 | - | 2047 | 2721 |
| 1994 | 1 | 408 | 304 | 0 | 1 | 327 | 866 | - | 1907 | 2693 |
| 1995 | 0 | 618 | 535 | 0 | 24 | 322 | 952 | - | 2451 | 3498 |
| 1996 | 0 | 462 | 460 | 0 | 22 | 156 | 944 | - | 2044 | 4054 |
| 1997 | 0 | 192 | 438 | 1 | 87 | 123 | 954 | - | 1795 | 3272 |
| 1998 | 0 | 172 | 433 | 0 | 111 | 65 | 841 | - | 1622 | 2705 |
| 1999 | 0 | 0 | 438 | 0 | 83 | 42 | 831 | - | 1394 | 2648 |
| 2000 | 0 | 135 | 417 | 0 | 98 | 20 | 754 | - | 1424 | 2247 |
| 2001 | 0 | 252 | 509 | 0 | 92 | 7 | 770 | - | 1630 | 2473 |
| 2002 | 0 | 79 | 280 | 0 | 89 | 14 | 643 | - | 1105 | 1828 |
| 2003 | 0 | 92 | 344 | 0 | 98 | 13 | 558 | - | 1105 | 1642 |
| 2004 | 0 | 50 | 278 | 0 | 45 | 17 | 469 | - | 859 | 1328 |
| 2005 | 0 | 48 | 156 | 0 | 69 | 10 | 269 | - | 552 | 561 |
| 2006 | 0 | 53 | 221 | 0 | 52 |  |  | 346 | 672 | 875 |
| 2007 | 0 | 104 | 191 | 0 | 5 |  |  | 667 | 967 | 1301 |
| 2008 | 0 | 92 | 172 | 0 | 149 |  |  | 874 | 1287 | 1545 |
| 2009 | 0 | 174 | 188 | 0 | 112 |  |  | 953 | 1427 | 1387 |
| 2010 | 0 | 271 | 318 | 0 | 288 |  |  | 822 | 1699 | 1698 |
| 2011 | 0 | 153 | 227 | 0 | 217 |  |  | 715 | 1312 | 1297 |
| 2012 | 0 | 140 | 214 | 0 | 142 |  |  | 590 | 1086 | 1132 |
| 2013 | 0 | 105 | 203 | 0 | 213 |  |  | 470 | 991 | 949 |
| 2014 | 0 | 126 | 246 | 0 | 57 |  |  | 465 | 894 | 948 |
| 2015 | 0 | 140 | 311 | 0 | 140 |  |  | 520 | 1110 | 1110 |
| 2016 | 0 | 189 | 408 | 0 | 146 |  |  | 694 | 1437 | 1437 |
| 2017 | 0 | 132 | 336 | 0 | 313 |  |  | 579 | 1359 | 1359 |
| 2018* | 0 | 119 | 301 | 0 | 289 |  |  | 683 | 1392 | 1392 |

* Preliminary. ** Historical landings data have been adjusted for area misreporting, mainly from Division 4.a to Division 6.a.

Table 13．3．Megrim in Subarea 4 and 2．a．Nominal catch（ $t$ ）of Megrim North Sea，as officially reported to ICES and WG best estimates of landings．

| 2 0 0 0 | $\begin{aligned} & \frac{\varepsilon}{工} \\ & \frac{10}{600} \\ & \stackrel{\sim}{\infty} \end{aligned}$ | $\begin{aligned} & \text { 늧 } \\ & \stackrel{y}{\pi} \\ & \frac{1}{\omega} \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { U } \\ & \stackrel{\text { Cu}}{\Gamma} \\ & \text { ㄴ } \end{aligned}$ |  |  |  |  | $\begin{aligned} & \text { त } \\ & \sum_{0}^{2} \\ & 0 \end{aligned}$ | $\begin{aligned} & \stackrel{\check{10}}{0} \\ & \text { in } \end{aligned}$ | $\begin{aligned} & \stackrel{\smile}{0} \\ & \frac{0}{0} \\ & z_{n} \end{aligned}$ |  |  |  | $\begin{aligned} & \text { 듣 } \\ & \stackrel{\pi}{4} \\ & \stackrel{\rightharpoonup}{u} \\ & \stackrel{1}{j} \\ & \stackrel{1}{2} \end{aligned}$ | $\stackrel{\checkmark}{〕}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 4 | 2 | － | － | 3 | － | 24 | － | － | － | 17 | － | － | 1126 | － | 1176 | 837 |
| 1991 | 3 | 1 | － | 6 | － | － | 28 | － | － | － | 9 | － | － | 1169 | － | 1216 | 878 |
| 1992 | 2 | 4 | 36 | 3 | － | － | 27 | － | － | － | 47 | － | － | 1372 | － | 1491 | 1025 |
| 1993 | 7 | 6 | 25 | 4 | － | － | 30 | － | － | － | 8 | － | － | 1736 | － | 1816 | 1081 |
| 1994 | 2 | 1 | 27 | 1 | － | － | 28 | － | － | － | 19 | － | － | 2000 | － | 2078 | 1207 |
| 1995 | 7 | 2 | 24 | 2 | － | － | 26 | － | － | － | 44 | － | － | 2193 | － | 2298 | 1172 |
| 1996 | 5 | 7 | 14 | 1 | － | － | 9 | － | － | － | 4 | － | － | 3221 | － | 3261 | 1199 |
| 1997 | 3 | 5 | 16 | 2 | － | － | 20 | － | － | － | 3 | － | － | 3091 | － | 3140 | 1584 |
| 1998 | 5 | 18 | 14 | 4 | － | － | 30 | － | － | － | 5 | － | － | 2628 | － | 2704 | 1548 |
| 1999 | 4 | 21 | ． | 1 | － | － | 26 | － | － | － | 4 | － | － | 2121 | － | 2177 | 1111 |
| 2000 | 10 | 29 | 7 | 3 | － | － | 20 | － | － | － | 2 | － | － | 2044 | － | 2115 | 1247 |
| 2001 | 2 | 52 | 5 | 1 | － | － | 11 | － | － | － | 2 | － | － | 1854 | － | 1927 | 1098 |
| 2002 | 5 | 8 | 6 | － | － | － | 9 | － | － | － | 3 | － | － | 1675 | － | 1706 | 975 |
| 2003 | 3 | 11 | 11 | 2 | － | 1 | 7 | ＜0．5 | － | － | 1 | － | － | 1235 | － | 1271 | 727 |
| 2004 | － | 7 | 9 | 2 | － | － | 11 | $<0.5$ | － | － | 1 | － | － | 1130 | － | 1160 | 739 |
| 2005 | － | 1 | 3 | 4 | － | － | 19 | $<0.5$ | － | － | 1 | － | － | 958 | － | 986 | n／a |
| 2006 | 0 | 3 | 4 | 1 |  | 0 | 6 | 1 | 0 | 0 |  |  |  |  | 1342 | 1357 | 1179 |
| 2007 | 0 | 11 | 18 | 4 |  | 0 | 1 | 1 | 0 | 0 |  |  |  |  | 1437 | 1472 | 1047 |
| 2008 | 0 | 31 | 20 | 1 |  | 0 | 1 | 4 | 0 | 0 |  |  |  |  | 1524 | 1581 | 1349 |
| 2009 | 0 | 54 | 9 | 0 |  | 0 | 0 | 6 | 0 | 0 |  |  |  |  | 1474 | 1543 | 1484 |
| 2010 | 0 | 22 | 1 | 0 |  | 0 | 1 | 2 | 0 | 0 |  |  |  |  | 1440 | 1466 | 1499 |
| 2011 | 0 | 23 | 10 | 3 |  | 0 | 0 | 1 | 0 | 0 |  |  |  |  | 1394 | 1431 | 1421 |
| 2012 | 0 | 35 | 5 | 3 |  | 0 | 0 | 1 | 0 | 0 |  |  |  |  | 1397 | 1441 | 1458 |
| 2013 | 0 | 48 | 7 | 3 |  | 0 | 0 | 17 | 0 | 0 |  |  |  |  | 1690 | 1765 | 1788 |
| 2014 | 0 | 35 | 7 | 1 |  | 0 | 0 | 12 | 0 | 0 |  |  |  |  | 1475 | 1530 | 1551 |
| 2015 | 0 | 26 | 1437 | 0 |  | 0 | 0 | 8 | 0 | 0 |  |  |  |  | 1175 | 1217 | 1230 |
| 2016 | 0 | 46 | 13 | 2 |  | 0 | 2 | 21 | 0 | 0 |  |  |  |  | 1278 | 1362 | 1361 |
| 2017 | 0 | 0 | 36 | 0 |  | 0 | 0 | 29 | 0 | 0 |  |  |  |  | 1171 | 1235 | 1235 |
| 2018＊ | 0 | 0 | 66 | 0 |  | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 1545 | 1611 | 1611 |

＊Preliminary．
${ }^{* *}$ Historical landings data have been adjusted for area misreporting，mainly from Division 4．a to Division 6．a．

Table 13.4 Time-series of megrim survey indices in ICES Area 6.a and Division 4 as used in the surplus production model.

| year | sco.6.a.q1 | sco.6.a.q4 | sco.4.a.q1 | sco.4.a.q3 | monk.6.a | monk.4.a |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 2.587277 | $N$ | $N$ | NA | NA | NA |  |
| 1986 | 1.687998 | $N$ | NA | 1.329749 | $N A$ | $N$ | NA |

Table 13.5. Lepidorhombus whiffiagonis in ICES areas 6.a and 4.a. Prior distributions on parameters.

| Parameter | Symbol | Prior distribution | Notes |
| :---: | :---: | :---: | :---: |
| Intrinsic rate of population growth | $r$ | Uniform( $0.001,2.0)$ |  |
| Carrying capacity | $K$ | $\text { Uniform }\left(\ln (\max (C)), \ln \left(10 \times \sum_{t=1985}^{2010} C_{t}\right)\right.$ | From the maximum catch to ten times the cumulative catch across all years assuming uniform distribution on the logarithmic scale |
| Catchabilities | $\log \left(q_{j}\right)$ | Uniform $(-11.0,0.0)$ | Uniformly distributed on log-scale. See catchability sensitivity in Section 2.2.3.1 |
| Process error variance | $\frac{1}{\sigma_{u}^{2}}$ | Gamma (shape $=0.001$, rate $=0.001)$ | Gamma distributed on inverse variance (precision) scale |
| Measurement error variances | $\frac{1}{\sigma_{z_{i} j}^{2}}$ | Gamma (shape $=0.001$, rate $=0.001)$ | Gamma distributed on inverse variance (precision) scale |
| Proportion of $K$ in 1985 | $a$ | Uniform( $0.01,2.0)$ |  |

Table 13.6. Parameter estimates for final assessment outputs.

| Parameter | Estimates <br> $\mathbf{2 0 1 3}$ | Estimates <br> $\mathbf{2 0 1 4}$ | Estimates <br> $\mathbf{2 0 1 5}$ | Estimates <br> $\mathbf{2 0 1 6}$ | Estimates <br> $\mathbf{2 0 1 7}$ | Estimates <br> $\mathbf{2 0 1 8}$ | Estimates <br> $\mathbf{2 0 1 9}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| r.hat | 0.67 | 0.55 | 0.51 | 0.51 | 0.507507 | 0.466 | 0.50 |
| M.hat | 39346 | 43134 | 47216 | 46840 | 42681 | 55129 | 44116 |
| $\mathrm{~F}_{\text {MSY }}$ | 6037 | 5660 | 5612 | 5362 | 5072 | 5362 | 5123 |
| $\mathrm{~B}_{\text {MSY }}$ | 19673 | 21567 | 23608 | 23420 | 21340 | 27565 | 22058 |
| B | 3624 | 4109 | 42416 | 42356 | 37610 | 38057 | 37062 |
| F | 0.09 | 0.08 | 0.07 | 0.07 | 0.07291 | 0.081 | 0.08 |
| $\mathrm{~B}_{\text {lim }}$ | 5902 | 6470 | 7082 | 7026 | 6402 | 8269 | 6617 |
| $\mathrm{~B}_{\text {trig }}$ | 9837 | 10783 | 11804 | 11710 | 10670 | 13782 | 11029 |

Table 13.7. Time-series of $B / B_{\text {msy }}$ and $F / F_{\text {msy }}$ estimates and landings and discards in tonnes for the final assessment.

| Year | $\mathrm{B} / \mathrm{B}_{\text {MSY }}$ | B/ $\mathrm{B}_{\text {MSY High }}$ | $B / B_{\text {MSY Low }}$ | Landings | Discards* | F/FMSY | F/FMSY High | F/F $\mathrm{MSY}^{\text {Low }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 2.093 | 3.699 | 0.438 | 4499 |  | 0.783 | 1.655 | 0.357 |
| 1986 | 1.514 | 2.286 | 0.399 | 2858 |  | 0.589 | 1.096 | 0.300 |
| 1987 | 1.414 | 2.075 | 0.383 | 4614 |  | 1.033 | 1.925 | 0.543 |
| 1988 | 1.344 | 2.119 | 0.369 | 5212 |  | 1.254 | 2.287 | 0.577 |
| 1989 | 1.077 | 1.609 | 0.326 | 3451 |  | 0.961 | 1.781 | 0.463 |
| 1990 | 0.979 | 1.430 | 0.309 | 3047 |  | 0.907 | 1.740 | 0.437 |
| 1991 | 0.916 | 1.330 | 0.318 | 3310 |  | 1.050 | 1.997 | 0.535 |
| 1992 | 0.978 | 1.430 | 0.330 | 3574 |  | 1.061 | 1.924 | 0.536 |
| 1993 | 1.058 | 1.555 | 0.338 | 3802 |  | 1.049 | 1.949 | 0.520 |
| 1994 | 1.169 | 1.855 | 0.354 | 3900 |  | 0.978 | 1.811 | 0.447 |
| 1995 | 1.189 | 1.863 | 0.353 | 4670 |  | 1.169 | 2.189 | 0.540 |
| 1996 | 1.148 | 1.841 | 0.338 | 5253 |  | 1.383 | 2.530 | 0.604 |
| 1997 | 0.958 | 1.421 | 0.307 | 4856 |  | 1.485 | 2.702 | 0.698 |
| 1998 | 0.920 | 1.423 | 0.277 | 4253 |  | 1.331 | 2.450 | 0.599 |
| 1999 | 0.890 | 1.387 | 0.265 | 3759 |  | 1.195 | 2.214 | 0.527 |
| 2000 | 0.827 | 1.286 | 0.249 | 3494 |  | 1.172 | 2.137 | 0.512 |
| 2001 | 0.765 | 1.181 | 0.237 | 3571 |  | 1.289 | 2.411 | 0.582 |
| 2002 | 0.775 | 1.195 | 0.241 | 2803 |  | 0.964 | 1.770 | 0.446 |
| 2003 | 0.808 | 1.330 | 0.236 | 2369 |  | 0.769 | 1.418 | 0.338 |
| 2004 | 0.808 | 1.240 | 0.245 | 2067 |  | 0.652 | 1.196 | 0.304 |
| 2005 | 0.796 | 1.168 | 0.237 | 1527 |  | 0.468 | 0.910 | 0.241 |
| 2006 | 0.898 | 1.289 | 0.263 | 2054 |  | 0.567 | 1.111 | 0.297 |
| 2007 | 1.021 | 1.473 | 0.312 | 2348 |  | 0.575 | 1.087 | 0.299 |
| 2008 | 1.131 | 1.619 | 0.330 | 2894 |  | 0.647 | 1.248 | 0.334 |
| 2009 | 1.245 | 1.807 | 0.381 | 2871 |  | 0.582 | 1.112 | 0.301 |
| 2010 | 1.265 | 1.757 | 0.392 | 3197 |  | 0.635 | 1.145 | 0.344 |
| 2011 | 1.302 | 1.849 | 0.426 | 3257 |  | 0.626 | 1.133 | 0.341 |
| 2012 | 1.422 | 2.060 | 0.466 | 2545 |  | 0.472 | 0.849 | 0.258 |
| 2013 | 1.617 | 2.491 | 0.526 | 2737 | 327 | 0.407 | 0.733 | 0.214 |
| 2014 | 1.624 | 2.346 | 0.519 | 2500 | 309 | 0.367 | 0.646 | 0.203 |
| 2015 | 1.506 | 2.094 | 0.495 | 2471 | 152 | 0.361 | 0.616 | 0.208 |
| 2016 | 1.608 | 2.261 | 0.586 | 2792 | 167 | 0.386 | 0.642 | 0.222 |
| 2017 | 1.782 | 2.709 | 0.664 | 2594 | 193 | 0.333 | 0.560 | 0.178 |
| 2018 | 1.681 | 2.198 | 1.154 | 3003 | 255 | 0.4 | 0.59 | 0.24 |

[^6]Table 13.8. Basis for the catch options.

| Variable | Value | Source | Notes |
| :--- | :---: | :--- | :--- |
| F (2018)/F | MSY | 0.402038 | ICES (2018a) | F (average 2015-2018) | B (UPDATE)/B MSY | 1.681026 | ICES (2018a) | Short-term forecast |
| :--- | :---: | :--- | :--- |
| Catch (2019) | 3258 | ICES (2018a) | Short-term forecast |
| Landings (2019) | 3003 | ICES (2018a) | Assuming discard rate of 7.98\%in total weight of catch (average <br> 2015-2017) |
| Discards (2019) | 255 | ICES (2018a) | Assuming discard rate of 7.98\% in total weight of catch (average <br> 2015-2017) |

Table 13.9. The management option table.

| Basis | Total catch (2020) | Wanted catch* (2020) | Unwanted catch* (2020) | Fishing mortality <br> $\mathrm{F}_{2019} / \mathrm{F}_{\mathrm{MSY}}$ | Stock size $\mathrm{B}_{2021} / \mathrm{B}_{\mathrm{MSY}}$ | Probability** of $\mathrm{Bi}-$ omass 2021 falling below MSY $\mathrm{B}_{\text {trigger }}$ | Probability** of Biomass 2021 falling below $\mathrm{B}_{\text {lim }}$ | $\% \text { B }$ <br> change <br> *** | \% TAC change ${ }^{\wedge}$ | \% Advice change ${ }^{\text {m }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |  |  |
| MSY <br> approach: $\mathrm{F}_{\mathrm{MSY}}$ | 8350 | 7684 | 666 | 1.00 | 1.42 | 0.057 | 0.01 | -15 | 4 | 1 |
| $\mathrm{F}_{\text {MSY Lower }}$ | 6450 | 5935 | 515 | 0.77 | 1.52 | 0.030 | 0.01 | -10 | -26 | -22 |
| $\mathrm{F}_{\text {MSY UPPER }}$ | 8350 | 7684 | 666 | 1.00 | 1.42 | 0.06 | 0.01 | -15 | 4 | 1 |
| Other options |  |  |  |  |  |  |  |  |  |  |
| $\mathrm{F}=0$ | 0 | 0 | 0 | 0.00 | 1.83 | 0.02 | 0.01 | 9 | -100 | -100 |
| $B(2021)=B_{\text {lim }}$ | 27350 | 25168 | 2182 | 3.27 | 0.50 | 0.50 | 0.30 | -70 | 215 | 232 |
| $B(2021)=B_{p a}$ | 16700 | 15368 | 1332 | 1.99 | 1.01 | 0.50 | 0.01 | -40 | 92.6 | 102.4 |
| $\begin{aligned} & \mathrm{B}(2021)=\mathrm{MSY} \\ & \mathrm{~B}_{\text {trigger }} \end{aligned}$ | 16700 | 15368 | 1332 | 1.99 | 1.01 | 0.50 | 0.01 | -40 | 92.6 | 102.4 |
| $F=F_{2018}$ | 3258 | 2998 | 260 | 0.40 | 1.68 | 0 | 0 | 0 | -62 | -61 |

* "Wanted" and "unwanted" catch are used to describe fish that would be landed and discarded in the absence of the EU landing obligation (for Division 6.a only), based on discard rate estimates for 2014-2017.
**Probabilities are based on bootstrap sampling and based on a two- and three-year projection of $F$ and $B$, respectively.
***B 2021 relative to B 2010.
${ }^{\wedge}$ Total catch in 2020 relative to TAC in 2019 ( 8669 t , which corresponds to the 2019 TAC for subareas 4 and 6).
${ }^{\wedge}$ Advice value 2020 relative to advice value 2019 ( 8250 t).


Figure 13.1. Maps of the northern continental shelf around the British Isles showing the biomass of megrim during the anglerfish surveys (SAMISS and IAMISS) 2005-2017.


Figure 13.2. Scottish IBTS Q3 4.a megrim biomass maps.


Figure 13.3. Scottish IBTS Q1 4.a megrim biomass maps.


Figure 13.4 Scottish IBTS Q1 4.a megrim biomass maps.


Figure 13.5. Scottish IBTS Q4 6.a megrim biomass maps.


Figure 13.6. Comparison of the delta-gamma cpue estimates for the two 6.a Scottish IBTS surveys using the full time-series or truncating the series to 2010 after which the survey design and ground gear was changed.


Figure 13.7. Megrim biomass estimates in ICES divisions 4, 6.a and 6.b from the anglerfish (AMISS) survey with $95 \%$ cls.


Figure 13.8. Megrim cpue estimates in ICES division 6.a Q1 top left panel and 6.a Q4.


Figure 13.9. Pearson residuals for the six survey indices.


Parameter value

Figure 13.10. Prior (red line) and posterior distributions (black line) for the parameters in the model.


Figure 13. 11. Time-series of catch and model estimates of total biomass and exploitation rate (median values are shown as solid lines and $95 \%$ confidence intervals shown as broken lines). The model fits to the various cpue series is also shown (observations dots, median fit solid line and $95 \%$ confidence intervals shown as broken lines).


Figure 13.12. Comparison of assessment results models fitted to a cpue generated using the full time-series of the $6 . a$ (red) and a truncated time-series (blue).


Figure 13.13. Comparison with previous assessments.


Figure 13.14. Kobe plot of stock status.

## 11 Megrim (Lepidorhombus spp.) in Division 6.b (Rockall)

## Type of assessment in 2019

The current assessment is based on survey trends in relative biomass from the ISP-Anglerfish survey conducted annually in 6.a, 4.a and 6.b.

## ICES advice applicable to 2018

Based on ICES approach to data-limited stocks, ICES advises that landings and catches should be no more than 339 t and 443 t respectively in 2018.

## ICES advice applicable to 2019

ICES advises that when the EU multiannual plan (MAP) is applied, catches in 2019 that correspond to the $F$ ranges in the plan are between 354 tonnes and 473 tonnes.

## General

## Stock description and management units

Megrim stock structure is uncertain. Data collected during an EC study contract (98/096) on the 'Distribution and biology of anglerfish and megrim in the waters to the west of Scotland,' showed significantly different growth parameters and significant population structure difference between megrim sampled in $6 . a$ and 6.b (Anon, 2001). Spawning fish occur in both areas but whether these populations are reproductively isolated is not clear. WKFLAT (2011) concluded that megrim in $6 . b$ should continue to be considered as a separate stock until further information is available.


TAC/Management area
[ A Assessment area

Management area (red box) and assessment area (blue hatched area).

The recent TACs are presented above in Section 5.3.1.1.

## Fishery in 2018

Ireland had the highest catches in 2018 followed by Scotland and Spain(Table 14.1). The majority of the landings and catches are from otter trawlers.

| Landings |  | Discards |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Nephrops trawls | Other Gears | Finfish trawls | Nephrops trawls | Other Gears | Finfish trawls |
| $0.00 \%$ | $8.00 \%$ | $92.00 \%$ | $0.00 \%$ | $0.00 \%$ | $100 \%$ |

## Data

As part of the 2011 benchmark, landings-at-age data was compiled from 1990 to 2010. However; available age data from $6 . b$ prior to 2002 was sparse. A common Subarea 6 ALK was applied to megrim from 6.a and 6.b, which allowed area-specific age data from the anglerfish survey to be collected from 2012.

## Landings

Official landings data for each country together with Working Group best estimates of landings from 6.b are shown in Table 14.1. The WG best estimates of landings are the same as the official statistics.

Catches of megrim comprise two species, Lepidorhombus whiffiagonis and L. boscii. Information available to the Working Group indicates that L. boscii, are a negligible proportion of the Scottish and Irish megrim catch (Kunzlik et al., 1995; Anon, 2001). It is not clear to the WG whether landings of other countries are accurately partitioned by megrim species. Megrim are caught in association with anglerfish by some fleets and are area-misreported along with anglerfish. However, it is unknown whether misreporting from Division 6.b is an issue.

## Discards

Discard data from Ireland and Scotland were available in InterCatch from 2016. The discard estimates for Scotland decreased in 2018. Discard data for 2014 were available for Ireland in InterCatch, but the estimate for Scotland based on discard rates in Area 6 were as reported to STECF and landings of 95 t . Total discard estimates were available from 2005-2018. To estimate catches prior to 2005, for the SPiCT analysis; a catch over landing ratio of 1.2 was used (derived from that observed ratio between 2015-2019). In 2018 discards represented approximately $21 \%$ of catch; decreasing from 233 to 203 tonnes (Figure 14.1, Table 14.2).

## Surveys

In 2005, Scotland initiated a new industry-science partnership survey to provide an absolute abundance estimate for anglerfish. Fourteen years of survey data are available and these cover the main distribution of the anglerfish fishery. The survey is also considered to have greater spatial coverage for megrim, and as such was recommended by WKAGME (2008) as the main source of data of megrim relative biomass, for all megrim stocks in the Northern Shelf.

The survey index for $6 . b$ is presented in Table 14.2. Biomass and abundance recovery have continued in 2019 after prior reduction in 2017. The stock has displayed a largely increasing abundance and biomass trend since 2005. The area-stratified survey provides a minimum estimate of absolute biomass; survey catches are raised based on swept area and weighted by area. The survey assumes that all megrim in the trawl path are retained e.g. $q=1$. Assuming full retention is overly optimistic, therefore the minimum estimate of stock biomass was provided.
The biomass dynamic model used in the Lez.27.4a6a assessment, provided megrim catchability estimates of $0.2-0.3$ for SAIMISS-Q2/IAMISS-Q2 6.a and 4 . a surveys. The upper q estimate of 0.3 is used in combination to scale the survey biomass estimate. This provides an absolute biomass
and catch estimate offering a relatively broad harvest ratio approximation of megrim in $6 . \mathrm{b}$ (Table 14.2). This indicates the harvest ratio for megrim ranges from 2 to $25 \%$ over the time-series; however in recent years, this value has typically been less than $10 \%$.

## Historical stock development

No analytical assessment has been agreed for this stock since 1999.

## State of the stock

The state of the stock is unknown.

## Short-term projections

There is no accepted analytical assessment for this stock.
Biological and MSY reference points

Precautionary approach reference points
No precautionary reference points have been defined for this stock.

## MSY evaluations

Proxy reference points (FMSY and $B_{\text {trigger }}$ ) were explored for the stock at WKProxy (ICES, 2016) and WGCSE 2016 (ICES, 2016). A biomass dynamic model (SPiCT-Stochastic Production model in Continuous Time) was used to explore these reference points. This analysis was updated again by WGCSE 2017 using the SPiCT r package (Pedersen and Berg, 2016). The summary plots are shown in Figure 14.4. The stochastic reference point estimates are shown below. These are not significantly different to the results obtained by WGCSE last year.

| Reference point | estimate | cilow | ciupp | est.in.log |
| :--- | :---: | :---: | :---: | :---: |
| $\mathrm{B}_{\mathrm{MSY}}$ | 3582 | 2170 | 5914 | 8.2 |
| $\mathrm{~F}_{\mathrm{MSY}}$ | 0.216 | 0.144 | 0.323 | -1.5 |
| MSYS | 774 | 460 | 1305 | 6.6 |

The general conclusion of WKProxy and WGCSE in 2016 is still valid; that the stock is currently exploited below F msy $_{\text {p }}$ proxy reference points and $B_{\text {MSY }}$ is above the proxy for MSY $B_{\text {trigger }}$.

## Yield-per-recruit analysis

It was not possible to define $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\text {max }}$ values for this stock due to the lack of international catch-at-age data and recent changes in fleet selectivity due to likely changes in targeting behaviour and recent changes in mesh selectivity, which, if fully implemented, will result in a significant change in age selectivity of the gear.

### 11.1 Uncertainties and bias in assessment and forecast

There is no accepted analytical assessment for this stock.

### 11.2 Recommendation for next Benchmark

This stock was subject to benchmark in 2011. WGCSE should review the available data, discuss assessment options and schedule a benchmark as soon as is practical.

## Management considerations

The TAC in 6 has not been fully utilised. However, the uptake rate is country specific, with full uptake being reported by some Member States. Partial quota by individual Member States may be an artefact of reduction in effort rather than reflective of a reduction in biomass. The TAC and assessment area are incompatible.

### 11.3 References

Kunzlik, P. A., A. W. Newton and A. W. Jermyn. 1995. Exploitation of monks (Lophius spp.) and megrims (Lepidorhombus spp.) by Scottish fishermen in ICES Division VIa (West of Scotland). Final report EU FAR contract MA-2-520.

Laurenson, C. and MacDonald, P. 2008. Collection of fisheries and biological data on megrim in ICES Subarea IVa. Scottish Industry Science Partnership Report No 05/08.

ICES. 2016. Report of the Workshop to consider MSY proxies for stocks in ICES category 3 and 4 stocks in Western Waters (WKProxy), 3-6 November 2015, ICES Headquarters, Copenhagen. ICES CM 2015/ACOM:61. 183 pp.
ICES. 2016a. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 4-13 May 2016, ICES Headquarters, Copenhagen, Denmark. ICES CM 2016/ACOM:13. 1031 pp.

Pedersen, M.W. and Berg, C.W. 2017. A stochastic surplus production model in continuous time. Fish Fish, 18: 226-243. doi:10.1111/faf. 12174.

Table 14.1. Megrim in Subarea 6.b. Nominal catch ( $t$ ) of Lez.27.6b, as officially reported to ICES and WG best estimates of landings (tonnes).

| $\begin{aligned} & \text { 㐫 } \\ & \end{aligned}$ | $\begin{aligned} & \underline{E} \\ & \stackrel{E}{D} \\ & \stackrel{D}{D} \\ & \infty \end{aligned}$ | $\begin{aligned} & \text { U } \\ & \text { 든 } \\ & \text { 꼬 } \end{aligned}$ | $\begin{aligned} & \text { ס } \\ & \underline{C} \\ & \underline{\underline{N}} \end{aligned}$ | $\begin{aligned} & \text { 듣 } \\ & \text { io } \end{aligned}$ |  |  |  | $\underset{ }{〕}$ | $\bar{\pi}$ <br> 0 <br> 0 <br> 0 <br>  <br>  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 |  |  | 240 | 587 | 14 |  | 204 |  | 1045 | 1045 |  |
| 1992 |  |  | 139 | 683 | 53 |  | 198 |  | 1073 | 1073 |  |
| 1993 |  |  | 128 | 594 | 56 |  | 147 |  | 925 | 925 |  |
| 1994 |  |  | 176 | 574 | 38 |  | 258 |  | 1046 | 1046 |  |
| 1995 |  |  | 117 | 520 | 27 |  | 152 |  | 816 | 816 |  |
| 1996 |  |  | 124 | 515 | 92 |  | 112 |  | 843 | 843 |  |
| 1997 |  |  | 141 | 628 | 76 |  | 164 |  | 1009 | 1009 |  |
| 1998 |  |  | 218 | 549 | 116 |  | 208 |  | 1091 | 1091 |  |
| 1999 |  |  | 127 | 404 | 57 |  | 278 |  | 866 | 866 |  |
| 2000 |  | 4 | 167 | 427 | 57 |  | 309 |  | 964 | 964 |  |
| 2001 |  | < 0.5 | 176 | 370 | 42 |  | 236 |  | 824 | 824 |  |
| 2002 |  | $<0.5$ | 87 | 120 | 41 |  | 207 |  | 455 | 455 |  |
| 2003 |  |  | 83 | 93 | 74 |  | 382 |  | 632 | 632 |  |
| 2004 |  |  | 43 | 71 | 42 |  | 372 |  | 528 | 528 |  |
| 2005 |  |  | 68 | 88 | 19 |  | 207 |  | 382 | 382 | 87 |
| 2006 |  |  | 95 | 59 | 9 |  | 181 |  | 344 | 344 | 75 |
| 2007 |  |  | 87 | 19 |  |  |  |  | 106 | 106 | 22 |
| 2008 |  |  | 68 | 84 |  | 1 | 141 |  | 294 | 294 | 59 |
| 2009 |  |  | 48 | 0 |  |  | 178 |  | 226 | 226 | 44 |
| 2010 |  |  | 47 | 0 |  |  |  | 92 | 139 | 139 | 26 |
| 2011 |  |  | 72 | 17 |  |  |  | 66 | 155 | 155 | 7 |
| 2012 |  |  | 120 | 15 |  |  |  | 89 | 224 | 224 | 21 |
| 2013 |  |  | 181 | 39 |  |  |  | 58 | 278 | 278 | 15 |
| 2014 |  |  | 230 | 18 |  |  |  | 95 | 343 | 343 | 15 |
| 2015 |  |  | 256 | 67 |  |  |  | 130 | 453 | 453 | 85 |
| 2016 |  |  | 272 | 27 |  |  |  | 106 | 405 | 405 | 145 |
| 2017 |  |  | 358 | 46 | 15 |  | 167 |  | 586 | 586 | 233 |
| 2018 |  |  | 438 | 62 | 14 |  | 249 |  | 263 | 763 | 203 |

Table14.2. Estimates of Lez.27.6b biomass and harvest ratio from SAMISS surveys.

| Year | Survey Biomass <br> (tonnes) | Survey q | Raised Biomass <br> (tonnes) | Landings <br> (tonnes) | Discards <br> (tonnes) | Catch <br> (tonnes) | Harvest Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 566 | 0.3 | 1886 | 382 | 87 | 469 | 0.25 |
| 2006 | 929 | 0.3 | 3098 | 344 | 75 | 419 | 0.14 |
| 2007 | 1267 | 0.3 | 4224 | 106 | 22 | 128 | 0.03 |
| 2008 | 1728 | 0.3 | 5759 | 294 | 59 | 353 | 0.06 |
| 2009 | 1605 | 0.3 | 5349 | 226 | 44 | 270 | 0.05 |
| 2010 | 1991 | 0.3 | 6636 | 139 | 26 | 165 | 0.02 |
| 2011 | 885 | 0.3 | 2949 | 155 | 7 | 162 | 0.05 |
| 2012 | 4320 | 0.3 | 14401 | 224 | 21 | 245 | 0.02 |
| 2013 | 3030 | 0.3 | 10101 | 278 | 15 | 293 | 0.03 |
| 2014 | 3318 | 0.3 | 11060 | 343 | 15 | 358 | 0.03 |
| 2015 | 3262 | 0.3 | 10872 | 453 | 85 | 538 | 0.05 |
| 2016 | 4507 | 0.3 | 15024 | 405 | 145 | 550 | 0.04 |
| 2017 | 3015 | 0.3 | 10067 | 586 | 233 | 819 | 0.08 |
| 2018 | 3984 | 0.3 | 13280 | 763 | 203 | 967 | 1.13 |
| 2019 | 4150 | 0.3 | 13835 | - | - | - | - |

Table14.3. SPICT results for Lez.27.6b.

Convergence: 0 MSG: relative convergence (4)
Objective function at optimum: 33.5910069
Euler time step (years): $1 / 16$ or 0.0625
Nobs C: 28, Nobs I1: 15

```
Priors
    logn ~ dnorm[log(2), 2^2]
logalpha ~ dnorm[log(1), 2^2]
logbeta ~ dnorm[log(1), 2^2]
Model parameter estimates w 95% CI
    estimate cilow ciupp log.est
alpha }7.3755203 0.9510703 5.719693e+01 1.9981664
beta 0.6823274 0.2629204 1.770767e+00 -0.3822456
r 0.6141472 0.0381455 9.887847e+00 -0.4875206
rc 0.4343541 0.2894908 6.517080e-01 -0.8338953
rold 0.3359916 0.0846066 1.334297e+00 -1.0906690
m 780.7226300 458.3001577 1.329975e+03 6.6602199
K 6348.4804684 3477.4518751 1.158987e+04 8.7559708
q 0.8318326 0.4372011 1.582671e+00 -0.1841240
n 
sdb 0.0470530 0.0063377 3.493384e-01 -3.0564800
sdf 0.3093946 0.1638362 5.842726e-01 -1.1731379
sdi 0.3470406 0.2347785 5.129820e-01 -1.0583136
sdc}\quad0.2111084 0.1299810 3.428713e-01 -1.5553835
```

Deterministic reference points (Drp)
estimate cilow ciupp log.est
Bmsyd 3594.8674782180 .02727285927 .9406028 .187262
Fmsyd $0.217177 \quad 0.1447454 \quad 0.325854-1.527043$
MSYd 780.722630 458.30015771329 .9751596 .660220
Stochastic reference points (Srp)
estimate cilow ciupp log.est rel.diff.Drp
Bmsys $3582.30101572170 .05974385913 .60703508 .183761-0.003507930$
Fmsys $0.2161803 \quad 0.1445486 \quad 0.3233096-1.531642-0.004610626$
MSYs 774.4104634459 .66707141304 .6650567 6.652102-0.008150931
States w 95\% CI (inp\$msytype: s)
estimate cilow ciupp log.est
B_2019.00 4401.24701312599 .69291667451 .25516438 .3896432
F_2019.00 $0.2014333 \quad 0.0942929 \quad 0.4303121-1.6022972$
B_2019.00/Bmsy $1.2286089 \quad 0.8477088 \quad 1.78065850 .2058826$
$\begin{array}{llllll}\text { F_2019.00/Fmsy } & 0.9317835 & 0.4116918 & 2.1089092 & -0.0706548\end{array}$



Figure 14.1. Lez.27.6b reported catch (landings and discards).


Figure 14.2 Lez.27.6b estimate biomass time-series.


Figure 14.3 Lez.27.6b estimate abundance time-series.


Figure 14.4. Lez.27.6b SPiCT model output. Top right: observed and fitted catch with 95 ci. Centre left: Biomass relative to $\mathrm{Bmsу}^{\text {. Centre: }} \mathrm{F}$ relative to $\mathrm{F}_{\text {msу. }}$. Corresponding MSY quantities are shown in each plot as horizontal lines ( 0.5 Bмsу in the case of the relative biomass plot). Centre right Kobe plot of stock trajectory.

## 12 Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 11 (West of Scotland, North Minch)

Nephrops stocks have previously been identified by WGNEPH on the basis of population distribution, and defined as separate Functional Units. The Functional Units (FU) in ICES Division 6.a (of which there are three) are defined by the groupings of ICES statistical rectangles given in Table 14.1 and illustrated in Figure 14.1. The functional unit is the level at which the WG collates fishery data (quantities landed and discarded, fishing effort and length distributions) and at which it performs assessments.

## Type of assessment in 2019

The assessment of North Minch Nephrops in 2019 is based on a combination of examining trends in fishery indicators and abundance estimated by underwater TV survey, both of which comprise an extensive dataseries for this FU. The assessment follows the process defined by the benchmark WG (WKNEPH 2009 and WKNEPH 2013). Further details on the assessment and catch options are provided in the stock annex.

## ICES advice applicable to 2018

'ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2014-2016, catches in 2018 should be no more than 2819 tonnes.

To ensure that the stock in functional unit (FU) 11 is exploited sustainably, management should be implemented at the functional unit level.'

## ICES advice applicable to 2019

'ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2015-2017, catches in 2019 should be no more than 3270 tonnes.

To ensure that the stock in functional unit (FU) 11 is exploited sustainably, management should be implemented at the functional unit level.'

### 12.1 General

Nominal landings as reported to ICES for divisions $6 . a$ and $6 . \mathrm{b}$ are presented in Table 14.1.1. Total official landings from Division 6.a were 8914 tonnes in 2018, mostly reported by the UK with only 65 tonnes reported from Ireland. Table 14.1.2 and Figure 14.1.1 shows WG estimates of landings in Division 6.a broken down by FU. Nephrops landings are also made from outside the functional units, from statistical rectangles where small pockets of suitable sediment exist, although these are generally small amounts. In 2018, 160 tonnes of landings were reported from outside the FUs which is lower than the long-term average (Table 14.1.2). The main areas of activity outside FUs are the Stanton Bank (to the west of the South Minch) and areas of suitable
sediment along the shelf edge and slope to the west of the Hebrides. There are no functional units in Division 6.b and only very small quantities of Nephrops are landed (Table 14.1.1(b)).

## Stock description and management units

TheNorth Minch (FU11) is located at the northern end of the west coast of Scotland (Figure14.1). Owing to its burrowing behaviour, the distribution of Nephrops is restricted to areas of mud, sandy mud and muddy sand. Within the North Minch functional unit these substrates are distributed according to prevailing hydrographic and bathymetric conditions. The area is characterised by numerous islands of varying size and sea lochs, which occur along the mainland coast. These topographical features create a diverse habitat with complex hydrography and a patchy distribution of soft sediments. Results from work on mapping the spatial extent of Nephrops habitat in the North Minch sea lochs indicate that the muddy habitat in these areas is only a very small proportion of the total Nephrops grounds in the North Minch (WKNEPH 2013).

## Management applicable to 2018 and 2019

The management unit is Subarea 6 and EU and international waters of 5.b. The TAC for this area is 15092 tonnes in 2019, up from 12129 tonnes in 2018.
Since 2016, fisheries catching Nephrops in Division 6.a have been covered by the EU landing obligation (EU, 2015a). Creel fisheries are exempt from the landing obligation due to high survivability of discards. Dermersal trawlers using a codend between $80 \mathrm{~mm}-110 \mathrm{~mm}$ and within 12 miles of shore are also exempt from the landings obligation.

## Ecosystem aspects

Details of the ecosystem aspects for this functional unit are provided in the stock annex if available.

## Fishery description

Information on developments in the fishery was provided by Marine Scotland compliance officers.

The fishery in 2018 followed a similar pattern to 2017. The fishery started slowly before a good summer fishery developed from May to August. The fishery the tailed off in September, similar to the situation in 2017. This was said to be a seasonal occurrence rather than being caused by bad weather. The majority of the Western Isles trawl fleet then relocated to the east coast and to the fishing grounds in the Firth of Forth/Eyemouth/Shields for the winter months. Trawls activity in the winter months was at a very low level. Activity in the Nephrops creel sector was slightly down on 2017, a high price for crab encouraged some vessels to target crab.

The largest part of the North Minch fleet is still based at Stornoway, numbering approximately 220 vessels in 2018. The majority of the Stornoway vessels (175) are below 10 m in length. Trawlers targeted Nephrops year round whereas some creelers targeted Nephrops in the winter and shifted to lobster and crab over the summer months.

The number of trawlers reduced at the end of 2017 and into 2018 due to vessels being sold to elsewhere in the UK and others licences were removed and sold on, leaving the vessels tied up.

The fleet were targeting the same areas in the North Minch as previous years. The notable changes were that the fleet went to the East coast from September on. This reduced activity meant
that static gear vessels, set Nephrops creels on area of the North Minch that would have, otherwise been trawled.

This area had the normal increase in fishing activity in the spring/summer caused by non-local visiting vessels from the east coast.

No major changes in gear use were reported. Since 2009, vessels have been required to fit 120 mm square meshed panels, in accordance with the west coast emergency measures (Council Reg. (EU) $43 / 2009$ ). Large SMPs ( 200 mm ) are also widely used in the North Minch and have been mandatory for all TR2 vessels with power $>112 \mathrm{~kW}$ fishing under the Scottish Conservation Credits scheme.

In 2018, there was a large reduction in landings and effort (Figure 14.2.1) in all three functional units on the west coast. This reduction was partly explained by the migration of the west coast fleet to the east coast to take advantage of improved Nephrops fishing opportunities in 4.a. Anecdotal information from the fishing industry suggests that an additional factor contributing to the migration of the fishing fleet was an issue with foreign crew being unable to work in the inshore grounds of the west coast therefore moved to the offshore grounds of the east coast. Further general information on the fishery can be found in the stock annex.

### 12.2 Data available

## InterCatch

Data for 2018 were successfully uploaded into InterCatch prior to the 2019 WG meeting. Uploaded data were worked up in InterCatch to generate 2018 raised international length-frequency distributions. Allocation schemes for any unsampled fleets are described in the stock annex. Data exploration in InterCatch has previously shown that outputs of raised data were very close to those generated by the previous method applied internally with differences being $<0.1 \%$. As such, InterCatch length-frequency outputs have been used in the stock assessment since 2012.

## Commercial catch

Official catch statistics (landings) reported to ICES are shown in Tables 15.1.1(a) and 15.1.1(b); these relate to the whole of 6 .a of which the North Minch is a part. Landings by gear category for FU11 provided by country have been reported since 1981 and are presented in Table 14.2.1. Landings from this fishery are usually only reported from Scotland; but between 2012 and 20142 tonnes of Nephrops were reported by Ireland and 1 tonne reported in 2017. Total reported Scottish landings in 2018 were 1961 tonnes, consisting of 1599 tonnes landed by trawlers targeting Nephrops ( $\sim 81 \%$ ), 329 tonnes landed by creel vessels ( $\sim 17 \%$ ) and 30tonnes by other trawlers. In 2018 there were no reports of Nephrops landed in the below minimum size (or more properly minimum conservation reference size) category in accordance with the EU landing obligation (EU, 2015b).

## Effort data

In 2015, WGCSE agreed that effort should be reported in kW days as this is likely to be more informative about changes in the actual fleet effort. Reported effort by Scottish trawlers targeting Nephrops (Métiers: OTB_CRU - Bottom Otter Trawls Targetting Crustaceans and OTT_CRU Multirig Otter Trawls Targeting Crustaceans) has shown a decreasing trend since 2000 (Figure 14.2.1) but in 2012 the effort increased by $20 \%$ due to the influx of vessels from the North Sea during the first quarter of the year. Since then, effort has declined although there was a small increase in 2016. In 2018, there was a $17 \%$ decline in effort from 2017 , which was attributed to
the majority of the North Minch fleet moving to the east coast in the winter months. Note that the year range in effort time-series (2000-2018) does not match with the more extensive year range available for landings, due to a lack of reliable effort data in the MSS in-house database. The effort is also slightly inconsistent with the landings data in that effort is provided for TR2 vessels while the 'Nephropstrawl' landings additionally includes landings bylarge mesh trawlers targeting Nephrops.

## Sampling levels

Length compositions of landings and discards are obtained during market and on-board observer sampling respectively. These sampling levels are shown in Table 14.2.2. Length compositions for the creel fishery are available for landings only as the small numbers of discardssurvive well and are not considered to be removed from the population. There was a decrease in sampling for this FU in 2018 due to a change in the sampling design that divides the discard sampling by east and west coast rather than by functional unit. This change was a trail and has since been reverted to the previous design due to the negative effect on sample numbers. This change had no effect on mean weights in landings and discards, these figures fell well within normal ranges.

## Length compositions

Figure 14.2.2 shows a series of annual length-frequency distributions for the period 2000 to 2018. Catch (removals) length compositions are shown for each sex along with the mean length for both. In both sexes the mean sizes fluctuate over time and has generally remained stable since 2012. This parameter might be expected to reduce in size if overexploitation were taking place. In 2018, the mean size was within the normal range of variation seen in this functional unit in recent years.

## Sex ratio

Males consistently make the largest contribution to the landings, although the proportion of males does vary between years (Figure 14.2.3(a)). This is likely due to the varying seasonal pattern in the fishery and associated relative catchability (due to different burrow emergence behaviour) of male and female Nephrops. Males are available throughout the year and the fishery is prosecuted in all quarters (although effort is usually reduced during the winter months when the weather is poor). Females are mainly taken in the summer when they emerge after egg hatching. The seasonal change in proportion of males to females is evident in Figure 14.2.3(b) in 2018 the normal trend where males dominate in quarters one and four but the ratio is more even (or often female dominated) in quarters two and three was not seen. Instead, males dominate the catch in quarters one, two and four however this is well within ranges observed in the past.

## Mean weights

The mean weight in the landings (trawls and creels combined) shows substantial interannual variation (Figure 14.2.4 and Table 14.2.3) decreasing between 2010 and 2012, followed by an increase in 2013-2015 and a decrease again in 2016 and stable in 2017 with a increase in 2018. Given the relatively larger size of creel caught Nephrops (compared to trawl) the proportion of creel landings has a substantial effect on overall size composition. The increases in mean weight to 2010 (and also size, Figure 14.2.2) in particular are due to a higher proportion of creel landings. Figure 14.2 .5 shows the mean weight by sample and gear type over the period 2009-2018. There is no obvious trend in North Minch trawl-caught mean weights, a slight increasing trend previously detected in these landings from 2009 to 2015 is no longer visible on inclusion of the 2016, 2017 and 2018 data. A decrease in the mean weight of creel caught males is still obvious, although this is largely driven by the cluster of high values from the start of 2010. The mean weight in the landings has a significant impact on the catch forecast. Due to the high interannual variability in
mean weights it was considered more appropriate to use a full time-series average, from 1999 (first year with creel and trawl length distributions combined) until 2018 for producing the catch options.

## Discarding

Discarding of undersized and unwanted Nephrops occurs in this fishery, and quarterly discard sampling has been conducted on the Scottish Nephrops trawler fleet since 1990. Discard rates fluctuate in this FU and averaged $\sim 8.6 \%$ by number in the last three years (Table 14.2.4). In 2018, the discard rate increased to $6.7 \%$ by number (from $5.2 \%$ in 2017).

It is likely that some Nephropssurvive the discarding process. An estimate of $25 \%$ (Charuauet al., 1982; Sangster et al., 1997; Wileman et al., 1999) survival is assumed for this FU in order to calculate removals (landings + dead discards) from the population. The discard survival rate for creel caught Nephrops has been shown to be high (ICES, 2013) and a value of $100 \%$ is used. The discard rate (adjusted for survival) which will be used in the provision of landings options for 2020 is $6.6 \%$ based on a three-year average of 2016-2018.

## Ab undance indices from UWTV surveys

Underwater TV surveys are available for this stock since 1994 (missing surveys in 1995and 1997). The stock area for this FU was updated in 2013 to $2908 \mathrm{~km}^{2}$ (see stock annex for further details).
In 2019, an error in the analysis of the 2018 TV survey for FU11 was noticed, this error was an incorrect camera angle. The analysis was redone with the corrected camera angle. The corrected abundance estimate for 2018 is 1188 million which is a change of $-2.22 \%$ from 1215 million original estimate. The corrected results are presented in this report.
In 2019, 47 valid stations were used in the survey final analysis (Table 14.2.5).Table 14.2.6 shows the basic analysis for the most recent TV survey conducted in FU11. At the 2012 SGNEPS meeting (ICES, 2012) it was decided that a CV (relative standard error) of $<20 \%$ was an acceptable precision level for UWTV survey estimates of abundance. The CV for the most recent TV survey was $10.5 \%$, lower than the precision level agreed (Table 14.2.6).

Figure 14.2.6 shows the distribution of stations in recent TV surveys (2014-2019), with the size of the symbols reflecting the Nephrops burrowdensity. Table14.2.5 and Figure 14.2 .7 show the timeseries estimated abundance for the TV surveys, with $95 \%$ confidence intervals on annual estimates.

The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009; ICES, 2013). A number of potential biases were highlighted including those due to edge effects, species burrow misidentification and burrow occupancy. The cumulative relative to absolute conversion factor estimated for FU11 was 1.33 meaning that the TV survey is likely to overestimate Nephrops abundance by $33 \%$.

### 12.3 Assessment

## Comparison with previous assessments

The assessment is the same as last year and is based on a combination of examining trends in fishery indicators and underwater TV abundance estimates. Landings predictions are derived by applying a harvest rate to the UWTV survey estimate of abundance and assuming a length composition derived from recent fishery data (including data from both trawl and creel fisheries).

No internal audit was conducted last year.

## State of the stock

The assessment summary is provided in Table 14.2.4. The underwater TV survey is presented as the best available information on the North Minch Nephrops stock. The surveys provide a fish-ery-independent estimate of Nephrops abundance. At present, it is not possible to extract any length or age-structure information from the survey and therefore it only provides information on abundance over the area of the survey.

TV survey estimated stock abundance in 2019 was 1232 million individuals, a $3.7 \%$ increase from the 2018 estimate. The stock is still well above the MSY $B_{\text {trigger }}$ value of 541 million, or the rounded value of 540 million individuals used in the provision of advice (Figure 14.2.7).

The calculated harvest ratein 2018 (dead removals/TV abundance $=6.4 \%$ ) is below the $\mathrm{F}_{\text {MSY }}$ proxy for this stock (the value associated with high long-term yield and low risk depletion) of $10.8 \%$.

### 12.4 Catch option table

Landings predictions at various harvest rates (based on principles established at WKNEPH (ICES, 2009)), including a selection of those equivalent to the per-recruit reference points, will be made on the basis of the 2019 UWTV survey conducted in June and presented in October 2019 for the provision of advice.

The table below shows the agreed inputs to the catch options table.

| Input | Data | 2019 assessment |
| :--- | :---: | :---: |
| Survey abundance (millions) | UWTV 2019 | 1232 |
| Mean weight in wanted catch (g) | $1999-2018$ | 25.9 |
| Mean weight in Unwatched catch (g) | average 2016-2018 | $1999-2018$ |
| Unwanted catch | Proportion by number | $8.6 \%$ |
| Discards survival | average 2016-2018 | $25 \%$ |
| Dead discard rate |  | $6.6 \%$ |

Due to the high interannual variability in mean weights it was considered more appropriate to use a full time-series average, from 1999 (first year with creel and trawl length distributions combined) until 2018 for producing the catch options.

### 12.5 Reference points

New reference point Fmsy were derived for this stock at WKMSYRef4 (ICES, 2016). This was updated on the basis of an average of estimated $\mathrm{F}_{\text {MSY }}$ proxy harvest rates over a period of years, this corresponds more closely to the methodology for finfish. In cases where there is a clear trend in the values a five-year average was chosen. Similarly, the five-year average of the F at $95 \%$ of the YPR obtained at the Fmsy proxy reference point was proposed as the Fmsy lower bound and the five-year average of the F above FMAx that leads to YPR of $95 \%$ of the maximum as the upper bound. Using an average value also has the advantage of reducing the effect of any unusually high or low estimates of the FMSY proxy, which occasionally appear. For this stock, the FMSY proxy has been revised from $10.9 \%$ to $10.8 \%$.

WKFMS YRef4 did not update the MSY Btrigger except for rounding to tens of millions. MSY Btrigger has been defined as the lowest stock size from which the abundance has increased (ICES, 2013) and is calculated as 541 million individuals and rounded to 540 million for use as MSY Btrigger in the advice. Full details are contained in the stock annex.

These reference points should remain under review by WGCSE and may be revised should improved data become available.
Table 14.2.4 and Figure 14.5.1 show the harvest rates for FU11. From 2006-2009 there was a sustained period of high, above FmsY proxy, harvest rates followed by two years of low harvest rates of around $6-7 \%$. There was a sudden increase in 2012, following this the harvest rate declined and has remained below the Fmsy proxy. In 2018, the harvest rate has decreased to $6.4 \%$. It is likely that prior to 2006, the estimated harvest rates may not be representative due to underreporting of landings.

### 12.6 Management strategies

Scotland has recently established a network of regional Inshore Fisheries Groups (rIFGs), nonstatutory bodies that aim to improve the management of Scotland's inshore fisheries out to six nautical miles, and to give commercial inshore fishermen a strong voice in wider marine management developments. The rIFGs will contribute to regional policies and initiatives relating to management and conservation of inshore fisheries, including impacts on the marine environment and the maintenance of sustainable fishing communities and measures designed to better conserve and sustainably exploit stocks of shellfish and sea fish (including salmon) in their local waters. Although no IFG proposals specific to the management of Nephrops fisheries have yet been adopted, some of the IFG management plans for the Scottish West Coast include spatial management of Nephrops fisheries and the introduction of creel limits.

On the 8th of February 2016, phase 1 of the fisheries management measures for inshore MPAs in Scottish waters came into force (SG, 2016). These measures relate to both NCMPA (Marine (Scotland) Act and the UK Marine and Coastal Access Act) andSACs (ECHabitats Directives - Council Directive 92/43/EEC) both of which have the aim of conserving biological diversity in Scottish waters and contribute toScotland'sMPA network (SG,2017a). Although not specific to the management of the Nephropsfishery, they will influence spatial patterns of fishing for Nephrops where controls on the two main gear types, demersal trawls and creels are implemented on Nephrops habitat. Within the North Minch functional unit, two MPAs are covered by fisheries management measures. Specifically the Wester Ross NCMPA where fishing activity is banned for dermersal gears for vessels over 500 kW in power and banned in certain areas for vessels below 500 kW . North of the main Nephrops ground is the Loch Laxford SAC where demersal trawling is banned (SG, 2016). The areas of the SAC and NCMPA relative to the estimated Nephrops habitat within the North Minch functional unit are displayed in Figure 14.6.1.

### 12.7 Quality of assessment and forecast

The length and sex composition of the landings data is considered to be well-sampled. Discard sampling has been conducted on a quarterly basis for Scottish Nephrops trawlers in this fishery since 1990, and is considered to represent the fishery adequately. The length compositions from 1999 onwards, are derived from both creel and trawl samples. The creel fishery accounted for greater than $20 \%$ of landings from 2009 to 2011, although this has decreased to $17 \%$ in 2018. This part of the fishery exhibits a length composition composed of larger animals.

There were concerns over the accuracy of historical landings and effort data prior to 2006 when Buyers and Sellers legislation was introduced and the reliability began to improve. Because of
this, the final assessment adopted is independent of historical landings data. Harvest rates since 2006 are also considered more reliable due to more accurate landings data reported under this legislation. Incorporation of creel length compositions (since the 2010 WG ) has also improved estimates of harvest rates. Underwater TV surveys have been conducted for this stock since 1994, with a continual annual series available since 1998. The number of valid stations in the survey has remained relatively stable throughout the time period. Confidence intervals around the abundance estimates are relatively small for this functional unit. In the provision of catch options based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. A three-year average (2016-2018) of discard rates (adjusted to account for some survival of discarded animals) has been used in the calculation of catch options.
The cumulative absolute conversion factor estimates for FU11 are largely based on expert opinion (see stock annex). The precision of these bias corrections cannot yet be characterised. The method to derive landings for the catch options is sensitive to the input dead discard rate and mean weight in landings and this introduces uncertainties in the catch forecasts. Precision estimates are needed for these forecast inputs.

The stock area was revised in 2013 (ICES, 2013) using integrated VMS-logbook data to more accurately estimate the spatial extent of Nephrops catches. Two other factors however, have the potential to increase the fished area further. Firstly, the inclusion of vessels smaller than 15 m would likely increase the fished area in some of the inshore locations and secondly, it is known that most of the sea lochs have areas of mud substrate and are typically fished by creel boats. In recent years, a number of TV surveys have taken place in the major North Minch sea lochs in an attempt to improve estimates of the ground area and Nephrops abundance. Work presented at the WKNEPH 2013 (ICES, 2013) showed that the total area of the sea lochs is $105 \mathrm{~km}^{2}$, which is considerably smaller than the offshore VMS area estimated to be $2908 \mathrm{~km}^{2}$. Therefore, it is unlikely that the exclusion of these inshore areas from the survey have an impact in the mean densities and overall abundance of Nephrops in the North Minch.

### 12.8 Recommendation for next benchmark

This stock was last benchmarked in 2013 (ICES, 2013). WGCSE will keep the stock under close review and recommend a future benchmark as required.

### 12.9 Management considerations

The WG, ACOM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level and management at the functional unit level could provide the controls to ensure that catch opportunities and effort were compatible and in line with the scale of the resource.

Creel fishing takes place in this area but overall effort by this fleet in terms of creel numbers is not known, and measures to control numbers are not in place. There is a need to ensure that the combined effort from all forms of fishing is taken into account when managing this stock.
There is a bycatch of other species in the area of the North Minch and STECF estimates that discards of whiting and haddock are high in 6.a generally. It is important that efforts are made to ensure that unwanted bycatch is kept to a minimum in this fishery. Efforts to reduce discards and unwanted bycatches of cod include the implementation of large square meshed panels (SMPs) of 120 mm under the west coast emergency measures, and SMPs of 200 mm which were introduced under the Scottish Conservation Credits scheme.

### 12.10 References

Council Reg. (EU) 43/2009.
Charuau A., Morizur Y., Rivoalen J.J. 1982. Survival of discarded Nephrops norvegicus in the Bay of Biscay and in the Celtic Sea, ICES-CM-1982/B:13.

Dobby H. 2009. Fmsy proxies for Nephrops stocks. Working document for WGNSSK, 5-11 May 2010 and WGCSE, 12-20 May, 2010.

ESRI. 2014. ArcGIS. Version 10.2.1. Environmental Systems Research Institute, Inc.: Redlands, CA.
EU. 2015a. COMMISSION DELEGATED REGULATION (EU) 2015/2438 of 12 October 2015 establishing a discard plan for certain demersal fisheries in north-western waters. Official Journal of the European Union, L 336/29. Available at: <http://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32015R2438\&from=EN $>$ [Accessed: 2016].

EU. 2015b. REGULATION (EU) 2015/812 OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL of 20 May 2015. Official Journal of the European Union, L 133/1. Available at: [http://eur-lex.eu-ropa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32015R0812\&from=EN](http://eur-lex.eu-ropa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32015R0812%5C&from=EN) [Accessed: 14/05/2017].
EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. Official Journal of the European Union, L 83: 1-17. http://data.europa.eu/eli/reg/2019/472/oj
ICES. 2010. Report of the Study Group on Nephrops Surveys (SGNEPS), 9-11 November 2010, Lisbon, Portugal. ICES CM 2010/SSGESST:22. 95 pp .
ICES. 2012. Report of the Study Group on Nephrops Surveys (SGNEPS), 6-8 March 2012, Acona, Italy. ICES CM 2012/SSGESST:19. 36 pp.

ICES. 2013 Report of the Benchmark Workshop on Nephrops assessment (WKNEPH). ICES CM 2013/45. 230 pp .
ICES. 2015. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4). ICES CM 2015/ACOM:58.185 pp.

ICES. 2016. EU request to ICES to provide Fmš ranges for selected stocks in ICES subareas 5 to 10. In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.4.1.

ICES. 2017. ICES Statistical Rectangles. International Council for the Exploration of the Sea: Copenhagen, Denmark. Available at: [http://geo.ices.dk/index.phpSource](http://geo.ices.dk/index.phpSource) [Accessed: 16/03/2017].

Sangster, G.I., Breen, M., Bova, D.J., Kynoch, R., O'Neill, F.G., Lowry. N., Moth-Poulsen, T. Hansen, U.J., Ulmestrand, M., Valentinsson, D., Hallback, H., Soldal, A.V., and Hoddevik, B. 1997. Nephropssurvival after escape and discard from commercial fishing gear. Presented at ICES FTFB Working Group, Hamburg, Germany 14-17 April, 1997, ICES CM 1997 CM/B.

SG. 2016. Simple guide to fisheries management measures in Marine Protected Areas. Marine Scotland (The Scottish Government): Edinburgh. Available at: [http://www.gov.scot/Resource/0049/00498320.pdf](http://www.gov.scot/Resource/0049/00498320.pdf) [Accessed: 16/05/2017].

SG. 2017a. Marine Protected Areas in Scotland's Seas - Guidelines on the selection of MPAs and development of the MPA network. Marine Scotland (The Scottish Government): Edinburgh. Available at: [http://www.gov.scot/Topics/marine/marine-environment/mpanetwork/mpaguidelines](http://www.gov.scot/Topics/marine/marine-environment/mpanetwork/mpaguidelines) [Accessed: 16/05/2017].

SG. 2017b. Marine conservation orders (MCOs) and fisheries management measures (MPAs and SACs) with effect May 2017. Marine Scotland (The Scottish Government): Edinburgh. Available at: [https://marinescotland.atkinsgeospatial.com/nmpi/default.aspx?availablelayers=838](https://marinescotland.atkinsgeospatial.com/nmpi/default.aspx?availablelayers=838) [Accessed: 31/05/2017].

SNH. 2015. Nature Conservation Marine Protected Areas. Scottish Natural Heritage (Scottish Government): Inverness. Available at: [https://gateway.snh.gov.uk/natural-spaces/index.jsp](https://gateway.snh.gov.uk/natural-spaces/index.jsp) [Accessed: 03/04/2017].

SNH. 2016. Special Areas of Conservation. Scottish Natural Heritage (Scottish Government): Inverness. Available at: [https://gateway.snh.gov.uk/natural-spaces/index.jsp](https://gateway.snh.gov.uk/natural-spaces/index.jsp) [Accessed: 03/04/2017].

Wessel, P. and Smith, W.H.F. 2016. GSHHG version 2.3.6-A Global Self-consistent, Hierarchical, Highresolution Geography Database. National Centers for Environmental Information, National Oceanic and Atmospheric Administration: Boulder, CO. [https://www.ngdc.noaa.gov/mgg/shorelines/gshhs.html](https://www.ngdc.noaa.gov/mgg/shorelines/gshhs.html) [Accessed: 31/05/2017].

Wileman, D.A., Sangster, G.I., Breen, M., Ulmestrand, M., Soldal, A.V. and Harris, R.R. 1999. Roundfish and Nephrops survival after escape from commercial fishing gear. EU Contract Final Report. EC Contract No: FAIR-CT95-0753.

Table 14.1. Nephrops functional units and descriptions by statistical rectangle.

| Functional Unit | Stock | Division | ICES Rectangles |
| :--- | :--- | :--- | :--- |
| 11 | North Minch | $6 . \mathrm{a}$ | $44-46$ E3-E4 |
| 12 | South Minch | $6 . \mathrm{a}$ | $41-43$ E2-E4 |
| 13 | Clyde | $6 . \mathrm{a}$ | $39-40$ E4-E5 |

Table 14.1.1(a). Nominal landings (tonnes) of Nephrops in Division 6.a, 1980-2018, as officially reported to ICES.

|  | France | Ireland | Spain | UK-(Engl+Wales+N.Irl) | UK- Scotland | UK | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 5 | 1 | - | - | 7422 | - | 7428 |
| 1981 | 5 | 26 | - | - | 9519 | - | 9550 |
| 1982 | 1 | 1 | - | 1 | 9000 | - | 9003 |
| 1983 | 1 | 1 | - | 11 | 10706 | - | 10719 |
| 1984 | 3 | 6 | - | 12 | 11778 | - | 11799 |
| 1985 | 1 | 1 | 28 | 9 | 12449 | - | 12488 |
| 1986 | 8 | 20 | 5 | 13 | 11283 | - | 11329 |
| 1987 | 6 | 128 | 11 | 15 | 11203 | - | 11363 |
| 1988 | 1 | 11 | 7 | 62 | 12649 | - | 12730 |
| 1989 | - | 9 | 2 | 25 | 10949 | - | 10985 |
| 1990 | - | 10 | 4 | 35 | 10042 | - | 10091 |
| 1991 | - | 1 | - | 37 | 10458 | - | 10496 |
| 1992 | - | 10 | - | 56 | 10783 | - | 10849 |
| 1993 | - | 7 | - | 191 | 11178 | - | 11376 |
| 1994 | 3 | 6 | - | 290 | 11047 | - | 11346 |
| 1995 | 4 | 9 | 3 | 346 | 12527 | - | 12889 |
| 1996 | - | 8 | 1 | 176 | 10929 | - | 11114 |
| 1997 | - | 5 | 15 | 133 | 11104 | - | 11257 |
| 1998 | - | 25 | 18 | 202 | 10949 | - | 11194 |
| 1999 | - | 136 | 40 | 256 | 11078 | - | 11510 |
| 2000 | 1 | 130 | 69 | 137 | 10667 | - | 11004 |
| 2001 | 9 | 115 | 30 | 139 | 10568 | - | 10861 |
| 2002 | - | 117 | 18 | 152 | 10225 | - | 10512 |
| 2003 | - | 145 | 12 | 81 | 10450 | - | 10688 |
| 2004 | - | 150 | 6 | 267 | 9941 | - | 10364 |
| 2005 | - | 153 | 17 | 153 | 7616 | - | 7939 |
| 2006 | - | 133 | 1 | 255 | 13419 | - | 13808 |


|  | France | Ireland | Spain | UK-(Engl+Wales+N.Irl) | UK- Scotland | UK | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | - | 155 | - | 2088 | 14120 | - | 16363 |
| 2008 | - | 56 | 1 | 419 | 14795 | - | 15271 |
| 2009 | - | 53 | - | 1226 | 11462 | - | 12741 |
| 2010 | - | 45 | 1 | 1962 | 10250 | - | 12258 |
| 2011 | - | 38 | - | 2517 | 10419 | - | 12974 |
| 2012 | - | 28 | - | 2502 | 11807 | - | 14337 |
| 2013* | - | 5 | - | - | - | 12866 | 12871 |
| 2014 | - | 51 | - | - | - | 12760 | 12811 |
| 2015** | - | 75 | - | - | - | 11653 | 11728 |
| 2016** | - | 107 | 0 | - | - | 14600 | 14707 |
| 2017 | - | 114 | - | - | - | 11442 | 11557 |
| 2018 | - | 65 | 0 | - | - | 8849 | 8914 |

Table 14.1.1(b). Nominal landings (tonnes) of Nephrops in Division 6.b, 1980-2018, as officially reported to ICES. There are no Functional Units in ICES Division $6 . b$ but occasional small landings are made.

|  | France | Germany | Ireland | Spain | UK-(Engl+Wales+N.Irl) | UK- Scotland | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | - | - | - | - | - | - | 0 |
| 1981 | - | - | - | - | - | - | 0 |
| 1982 | - | - | - | - | - | - | 0 |
| 1983 | - | - | - | - | - | - | 0 |
| 1984 | - | - | - | - | - | - | 0 |
| 1985 | - | - | - | - | - | - | 0 |
| 1986 | - | - | - | 8 | - | - | 8 |
| 1987 | - | - | - | 18 | 11 | - | 29 |
| 1988 | - | - | - | 27 | 4 | - | 31 |
| 1989 | - | - | - | 14 | - | - | 14 |
| 1990 | - | - | - | 10 | 1 | - | 11 |
| 1991 | - | - | - | 30 | - | - | 30 |
| 1992 | - | - | - | 2 | 4 | 1 | 7 |
| 1993 | - | - | - | 2 | 6 | 9 | 17 |
| 1994 | - | - | - | 5 | 16 | 5 | 26 |
| 1995 | 1 | - | - | 2 | 26 | 1 | 30 |
| 1996 | - | 6 | - | 5 | 65 | 5 | 81 |
| 1997 | - | - | 1 | 3 | 88 | 23 | 115 |
| 1998 | - | - | 1 | 6 | 46 | 7 | 60 |
| 1999 | - | - | - | 5 | 2 | 5 | 12 |
| 2000 | 2 | - | 8 | 3 | 4 | 4 | 21 |
| 2001 | 1 | - | 1 | 14 | 2 | 7 | 25 |
| 2002 | 1 | - | - | 7 | 3 | 7 | 18 |
| 2003 | - | - | 1 | 5 | 6 | 18 | 30 |
| 2004 | - | - | - | 2 | 7 | 13 | 22 |
| 2005 | 3 | - | 1 | 1 | 5 | 7 | 17 |
| 2006 | - | - | - | - | 1 | 3 | 4 |


|  | France | Germany | Ireland | Spain | UK-(Engl+Wales+N.Irl) | UK- Scotland | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | - | - | - | 2 | 3 | - | 5 |
| 2008 | - | - | - | - | - | - | 0 |
| 2009 | - | - | - | - | - | - | 0 |
| 2010 | - | - | - | - | - | - | 0 |
| 2011 | - | - | - | - | - | - | 0 |
| 2012 | - | - | - | - | - | - | 0 |
| 2013 | - | - | - | - | - | - | 0 |
| 2014 | - | - | - | - | - | - | 0 |
| 2015 | - | - | - | - | - | - | 0 |
| 2016 | - | - | - | - | - | 0 | 0 |
| 2017 | - | - | - | - | - | 2 | 2 |
| 2018 | - | - | - | - | - | 0 | 0 |

Table 14.1.2. Nephrops, Total Nephrops landings (tonnes) by Functional Unit plus Other rectangles, 1981-2018.

| Year | FU11 | FU12 | FU13 | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 2861 | 3652 | 2968 | 39 | 9520 |
| 1982 | 2799 | 3552 | 2620 | 27 | 8998 |
| 1983 | 3197 | 3413 | 4076 | 34 | 10720 |
| 1984 | 4143 | 4300 | 3310 | 36 | 11789 |
| 1985 | 4060 | 4008 | 4286 | 104 | 12458 |
| 1986 | 3381 | 3484 | 4341 | 89 | 11295 |
| 1987 | 4084 | 3892 | 3009 | 257 | 11242 |
| 1988 | 4035 | 4473 | 3664 | 529 | 12701 |
| 1989 | 3205 | 4745 | 2812 | 212 | 10974 |
| 1990 | 2546 | 4430 | 2909 | 182 | 10067 |
| 1991 | 2793 | 4442 | 3038 | 255 | 10528 |
| 1992 | 3559 | 4237 | 2803 | 248 | 10847 |
| 1993 | 3193 | 4458 | 3343 | 344 | 11338 |
| 1994 | 3614 | 4414 | 2630 | 441 | 11099 |
| 1995 | 3655 | 4682 | 3987 | 460 | 12784 |
| 1996 | 2872 | 3995 | 4057 | 239 | 11163 |
| 1997 | 3046 | 4344 | 3621 | 243 | 11254 |
| 1998 | 2441 | 3730 | 4841 | 157 | 11169 |
| 1999 | 3257 | 4052 | 3752 | 438 | 11499 |
| 2000 | 3247 | 3953 | 3417 | 421 | 11038 |
| 2001 | 3259 | 3991 | 3182 | 420 | 10852 |
| 2002 | 3440 | 3305 | 3384 | 397 | 10526 |
| 2003 | 3269 | 3879 | 3173 | 433 | 10754 |
| 2004 | 3082 | 3869 | 2973 | 403 | 10327 |
| 2005 | 2949 | 3848 | 3395 | 254 | 10446 |
| 2006 | 4166 | 4633 | 4780 | 241 | 13820 |
| 2007 | 3978 | 5471 | 6660 | 420 | 16529 |
| 2008 | 3799 | 5356 | 5923 | 128 | 15206 |
| 2009 | 3496 | 4285 | 4779 | 185 | 12745 |


| Year | FU11 | FU12 | FU13 | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | 2413 | 3846 | 5843 | 569 | 12671 |
| 2011 | 2697 | 3702 | 6432 | 219 | 13050 |
| 2012 | 3542 | 3989 | 6687 | 435 | 14653 |
| 2013 | 3413 | 3776 | 5435 | 234 | 12858 |
| 2014 | 3257 | 3179 | 6207 | 53 | 12696 |
| 2015 | 3002 | 3400 | 5147 | 309 | 11858 |
| 2016 | 3529.4 | 4402 | 6447 | 236 | 14614.4 |
| 2017 | 2448 | 3652 | 5222 | 250 | 11572 |
| 2018 | 1961 | 2536 | 4141 | 160 | 8798 |

*Includes below minimum size landed discards of 0.4 t .

Table 14.2.1. Nephrops, North Minch (FU11), Nominal Landings of Nephrops, 1981-2018.

| UK Scotland |  |  |  |  |  | Other United Kingdom and Ireland | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | Nephrops trawl | other | creel | Below Minimum Size | Subtotal |  |  |
| 1981 | 2320 | 171 | 370 | 0 | 2861 | 0 | 2861 |
| 1982 | 2323 | 105 | 371 | 0 | 2799 | 0 | 2799 |
| 1983 | 2784 | 96 | 317 | 0 | 3197 | 0 | 3197 |
| 1984 | 3449 | 160 | 534 | 0 | 4143 | 0 | 4143 |
| 1985 | 3235 | 117 | 708 | 0 | 4060 | 0 | 4060 |
| 1986 | 2641 | 203 | 537 | 0 | 3381 | 0 | 3381 |
| 1987 | 3459 | 143 | 482 | 0 | 4084 | 0 | 4084 |
| 1988 | 3450 | 148 | 437 | 0 | 4035 | 0 | 4035 |
| 1989 | 2603 | 112 | 490 | 0 | 3205 | 0 | 3205 |
| 1990 | 1941 | 134 | 471 | 0 | 2546 | 0 | 2546 |
| 1991 | 2229 | 126 | 438 | 0 | 2793 | 0 | 2793 |
| 1992 | 2978 | 149 | 432 | 0 | 3559 | 0 | 3559 |
| 1993 | 2699 | 86 | 408 | 0 | 3193 | 0 | 3193 |
| 1994 | 2916 | 245 | 453 | 0 | 3614 | 0 | 3614 |
| 1995 | 2940 | 183 | 532 | 0 | 3655 | 0 | 3655 |
| 1996 | 2354 | 148 | 370 | 0 | 2872 | 0 | 2872 |
| 1997 | 2553 | 102 | 391 | 0 | 3046 | 0 | 3046 |
| 1998 | 2023 | 68 | 350 | 0 | 2441 | 0 | 2441 |
| 1999 | 2792 | 56 | 409 | 0 | 3257 | 0 | 3257 |
| 2000 | 2695 | 28 | 524 | 0 | 3247 | 0 | 3247 |
| 2001 | 2649 | 42 | 568 | 0 | 3259 | 0 | 3259 |
| 2002 | 2775 | 79 | 586 | 0 | 3440 | 0 | 3440 |
| 2003 | 2606 | 45 | 618 | 0 | 3269 | 0 | 3269 |
| 2004 | 2391 | 30 | 661 | 0 | 3082 | 0 | 3082 |
| 2005 | 2270 | 23 | 656 | 0 | 2949 | 0 | 2949 |
| 2006 | 3446 | 23 | 697 | 0 | 4166 | 0 | 4166 |


| UK Scotland |  |  |  |  |  | Other United Kingdom and Ireland | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | Nephrops trawl | other | creel | Below Minimum Size | Subtotal |  |  |
| 2007 | 3361 | 26 | 591 | 0 | 3978 | 0 | 3978 |
| 2008 | 3229 | 13 | 557 | 0 | 3799 | 0 | 3799 |
| 2009 | 2849 | 34 | 613 | 0 | 3496 | 0 | 3496 |
| 2010 | 1783 | 9 | 621 | 0 | 2413 | 0 | 2413 |
| 2011 | 2109 | 17 | 571 | 0 | 2697 | 0 | 2697 |
| 2012 | 2963 | 12 | 565 | 0 | 3540 | 2 | 3542 |
| 2013 | 2356 | 480 | 575 | 0 | 3411 | 2 | 3413 |
| 2014 | 2752 | 13 | 490 | 0 | 3255 | 2 | 3257 |
| 2015 | 2561 | 23 | 418 | 0 | 3002 | 0 | 3002 |
| 2016 | 3039 | 15 | 475 | 0.4 | 3529.4 | 0 | 3529.4 |
| 2017 | 2041 | 45 | 361 | 0 | 2447 | 1 | 2448 |
| 2018 | 1599 | 30 | 329 | 0 | 1958 | 3 | 1961 |

*Below minimum size landings not rounded to show it was reported.

Table 14.2.2. Nephrops Scottish sampling levels all FUs in 6.a (including N. Irish for Cly de).
*Number of trips expressed as number of hauls for discards.

|  |  | 2016 |  | 2017 |  | 2018 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FU |  | N trips* | $N$ measured | N trips* | $N$ measured | N trips* | $N$ measured |
| North Minch | Landings | 44 | 32483 | 48 | 33118 | 44 | 48 |
|  | Discards | 23 | 4402 | 53 | 5800 | 30082 | 4136 |
| South Minch | Landings | 37 | 20439 | 60 | 40131 | 36 | 29 |
|  | Discards | 8 | 1274 | 53 | 5164 | 22837 | 2547 |
| Clyde | Landings | 22 | 19069 | 27 | 21769 | 19 | 14517 |
|  | N.Irish Landings | 28 | 18218 |  |  | 4 | 2019 |
|  | Discards | 21 | 3337 | 39 | 3332 | 15 | 1753 |

Table 14.2.3. Nephrops mean weight in the landings (FU11-13).

| Year | FU11 | FU12 | FU13 |
| :---: | :---: | :---: | :---: |
| 1990 | 21.39 | 19.99 | 24.27 |
| 1991 | 25.35 | 21.74 | 20.65 |
| 1992 | 21.66 | 24.10 | 25.16 |
| 1993 | 20.79 | 21.26 | 29.44 |
| 1994 | 23.45 | 24.96 | 25.28 |
| 1995 | 22.24 | 21.96 | 19.24 |
| 1996 | 26.68 | 23.10 | 21.68 |
| 1997 | 21.71 | 23.37 | 24.21 |
| 1998 | 23.65 | 22.18 | 17.98 |
| 1999* | 22.70 | 25.14 | 17.39 |
| 2000 | 24.19 | 27.30 | 19.96 |
| 2001 | 25.33 | 23.79 | 19.46 |
| 2002 | 25.93 | 26.83 | 16.35 |
| 2003 | 26.03 | 27.86 | 19.13 |
| 2004 | 25.16 | 27.37 | 18.80 |
| 2005 | 27.65 | 28.11 | 17.96 |
| 2006 | 24.52 | 26.24 | 19.27 |
| 2007 | 23.61 | 23.95 | 19.05 |
| 2008 | 23.90 | 23.91 | 16.59 |
| 2009 | 25.42 | 23.87 | 18.31 |
| 2010 | 29.39 | 25.86 | 21.21 |
| 2011 | 27.56 | 31.10 | 19.34 |
| 2012 | 23.43 | 29.17 | 21.83 |
| 2013 | 27.52 | 27.48 | 20.72 |
| 2014 | 27.96 | 29.91 | 20.79 |
| 2015 | 28.74 | 28.15 | 22.21 |
| 2016 | 25.76 | 24.76 | 17.70 |
| 2017 | 25.89 | 27.76 | 17.02 |
| 2018 | 27.39 | 27.27 | 16.14 |
| Average** | 25.90 | 26.79 | 16.95 |

[^7]** Average for FU11 and FU12 (1999-2018); FU13 (2015-2018).

Table 14.2.4. Nephrops, North Minch (FU11): Adjusted TV survey abundance, landings, discard rate (proportion by number) and estimated harvest rate.

| YEAR | LANDINGS IN NUMBERS (MILLIONS) | DISCARDS IN NUMBERS (MILLIONS) | REMOVALS IN NUMBERS (MILLIONS)** | ADJUSTED SURVEY VMS (MILLIONS)* | HARVEST RATE VMS | LANDINGS (TONNES) | DISCARDS (TONNES) | DISCARD RATE | DEAD DISCARD RATE | MEAN WEIGHT IN LANDINGS (g) | MEAN WEIGHT IN DISCARDS (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 144 | 28 | 165 | 794 | 20.7 | 3257 | 273 | 16.4 | 12.8 | 22.7 | 9.69 |
| 2000 | 134 | 10 | 142 | 1166 | 12.1 | 3247 | 100 | 6.9 | 5.2 | 24.19 | 10.08 |
| 2001 | 129 | 17 | 141 | 1092 | 13 | 3259 | 160 | 11.7 | 9.1 | 25.33 | 9.32 |
| 2002 | 133 | 28 | 154 | 1337 | 11.5 | 3440 | 277 | 17.6 | 13.8 | 25.93 | 9.78 |
| 2003 | 126 | 30 | 148 | 1751 | 8.5 | 3269 | 299 | 19.2 | 15.2 | 26.03 | 10 |
| 2004 | 122 | 18 | 136 | 1751 | 7.8 | 3082 | 202 | 13 | 10.1 | 25.16 | 11.02 |
| 2005 | 107 | 50 | 144 | 1540 | 9.4 | 2949 | 507 | 32 | 26.1 | 27.65 | 10.09 |
| 2006 | 170 | 74 | 225 | 1762 | 12.8 | 4166 | 757 | 30.3 | 24.6 | 24.52 | 10.27 |
| 2007 | 168 | 12 | 177 | 1206 | 14.7 | 3978 | 214 | 6.5 | 5 | 23.61 | 18.1 |
| 2008 | 159 | 19 | 173 | 1047 | 16.5 | 3799 | 194 | 10.5 | 8.1 | 23.9 | 10.36 |
| 2009 | 138 | 35 | 164 | 1195 | 13.7 | 3496 | 327 | 20.3 | 16 | 25.42 | 9.34 |
| 2010 | 82 | 12 | 91 | 1293 | 7 | 2413 | 128 | 12.4 | 9.6 | 29.39 | 10.98 |
| 2011 | 96 | 16 | 108 | 1726 | 6.3 | 2697 | 154 | 14.2 | 11 | 27.56 | 9.66 |
| 2012 | 151 | 21 | 167 | 891 | 18.7 | 3542 | 213 | 12.6 | 9.3 | 23.43 | 10.33 |
| 2013 | 122 | 24 | 140 | 1403 | 10 | 3413 | 364 | 16.4 | 12.8 | 27.52 | 15.18 |
| 2014 | 115 | 8 | 121 | 1251 | 9.6 | 3257 | 77 | 6.3 | 4.8 | 27.96 | 9.99 |
| 2015 | 103 | 15 | 114 | 1445 | 7.9 | 3002 | 143 | 12.6 | 9.8 | 28.74 | 9.66 |
| 2016 | 136 | 22 | 152 | 1422 | 10.7 | 3529*** | 266 | 14 | 10.9 | 25.76 | 12.05 |
| 2017 | 93 | 5 | 97 | 1050 | 9.3 | 2448 | 64 | 5.2 | 4 | 25.89 | 12.51 |
| 2018 | 72 | 5 | 76 | 1188 | 6.4 | 1961 | 59 | 6.7 | 5.1 | 27.39 | 11.46 |
| Average**** |  |  |  |  |  |  |  |  | 6.67 | 25.90 | 10.99 |

${ }^{*}$ harvest rates previous to 2006 are unreliable.
** Removals numbers take the dead discard rate into account.
*** Includes 0.4 tonnes of below minimum size landings.
**** Dead discard average: 2016-2018; Mean weight in landings and discards average: 1999-2018.

Table 14.2.5. Nephrops, North Minch (FU11): Results of the 1994-2018 TV surveys (values adjusted for bias).

| YEARS | NUMBER OF VALID STATIONS | MEAN DENSITY (BURROWS/M ${ }^{2}$ ) | ABUNDANCE (SEDIMENT; MILLIONS) | 95\% CONFIDENCE INTERVAL (SEDIMENT; MILLIONS) | ABUNDANCE (VMS; MILLIONS) | 95\% CONFIDENCE INTERVAL (VMS; MILLIONS) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 41 | 0.29 | 500 | 74 | 820 | 122 |
| 1995 |  |  |  | No Survey |  |  |
| 1996 | 38 | 0.19 | 330 | 47 | 541 | 76 |
| 1997 |  |  |  | No Survey |  |  |
| 1998 | 38 | 0.31 | 547 | 77 | 898 | 127 |
| 1999 | 36 | 0.27 | 484 | 89 | 794 | 147 |
| 2000 | 39 | 0.40 | 711 | 82 | 1166 | 134 |
| 2001 | 56 | 0.38 | 666 | 81 | 1092 | 133 |
| 2002 | 37 | 0.46 | 815 | 91 | 1337 | 149 |
| 2003 | 41 | 0.60 | 1068 | 129 | 1751 | 211 |
| 2004 | 38 | 0.60 | 1068 | 107 | 1751 | 175 |
| 2005 | 41 | 0.53 | 939 | 100 | 1540 | 164 |
| 2006 | 30 | 0.61 | 1074 | 101 | 1762 | 165 |
| 2007 | 36 | 0.41 | 735 | 92 | 1206 | 150 |
| 2008 | 41 | 0.36 | 638 | 95 | 1047 | 157 |
| 2009 | 26 | 0.41 | 729 | 138 | 1195 | 227 |
| 2010 | 37 | 0.44 | - | - | 1293 | 231 |
| 2011 | 41 | 0.59 | - | - | 1726 | 226 |
| 2012 | 41 | 0.31 | - | - | 891 | 181 |
| 2013 | 41 | 0.48 | - | - | 1403 | 206 |
| 2014 | 44 | 0.43 | - | - | 1251 | 171 |
| 2015 | 41 | 0.50 | - | - | 1445 | 370 |
| 2016 | 39 | 0.49 | - | - | 1422 | 290 |
| 2017 | 42 | 0.36 | - | - | 1050 | 149 |
| 2018 | 44 | 0.40 | - | - | 1188 | 244 |
| 2019 | 47 | 0.42 | - | - | 1232 | 256 |

Table 14.2.6. Nephrops, North Minch (FU11): Results of the 2019 TV survey.

| STRATUM | AREA <br> $\left(\mathrm{km}^{2}\right)$ | NUMBER <br> OF STA- <br> TIONS | MEAN <br> BURROW <br> DENSITY <br> $\left(\mathrm{no}. / \mathrm{m}^{2}\right)$ | OB- <br> SERVED <br> VARI- <br> ANCE | ABUN- <br> DANCE <br> (MILLIONS) | STRATUM <br> VARI- <br> ANCE | PROPORTION <br> OF TOTAL <br> VARIANCE | SURVEY <br> PRECISION <br> LEVEL <br> (RSE) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2019 TV survey |  |  |  |  |  |  |  |  |
| VMS | 2908 | 47 | 0.423 | 0.092 | 1231.6 | 16438 | 1 |  |
| Total | 2908 | 47 |  |  | 1231.6 | 16438 | 1 | 0.105 |



Figure 14.1. Nephrops Functional Units in 6.a. North Minch (FU11), South Minch (FU12), Cly de (FU13).


Figure 14.1.1. Nephrops in Division 6.a. Landings (tonnes) by functional unit (FU11 to 13) and from rectangles outside the functional units (Other).

## Landings - International



Effort - Scottish Nephrops trawlers


Figure 14.2.1. Nephrops, North Minch (FU11). Long-term landings and effort.

## Length frequencies for catch (dotted) and landed(solid): Nephrops in FU11



Figure 14.2.2. Nephrops, North Minch (FU11), Catch length-frequency distribution and mean sizes (red line) for Nephrops in the North Minch, 2000-2018.


Figure 14.2.3 (a). Nephrops, North Minch (FU11), Landings by quarter and sex from Scottish trawlers.

NM_NA_NM Male sex ratio


Figure 14.2.3 (b). Nephrops, North Minch (FU11), Proportion of males by quarter (1980-2018).


Figure 14.2.4. Nephrops, (FU11 North Minch, FU12 South Minch and FU13 Cly de), mean weight in the landings from 1990-2018 (from Scottish market sampling data).


Figure 14.2.5. Nephrops, (FU11 North Minch, FU12 South Minch, FU13 Cly de), mean weight in landings 20092018 by sample date, sex, métier and functional unit.


Figure 14.2.6. Nephrops, North Minch (FU11), TV survey station distribution and relative density (burrows/m²), 2014-2019. Bubbles in these figures are all scaled the same. Crosses represent zero observations.

## North Minch



Figure 14.2.7. Nephrops, North Minch (FU11), time-series of revised TV survey abundance estimates (adjusted for bias), with $95 \%$ confidence intervals, 1994-2017 (no survey in 1995 and 1997). The dashed blue line is the rounded $\mathrm{B}_{\text {trigger }}$ value of 540 million individuals.


Figure 14.5.1. Nephrops, North Minch (FU11), harvest rate, 1995-2017 (no survey data in 1995 and 1997). The blue dashed and solid lines are the Fmsy proxy harvest rate ( $\mathbf{1 0 . 8 \%}$ ) and the harvest rate respectively. Harvest rates prior to 2006 are unreliable.


Figure 14.6.1. The area of Nephrops habitat (estimated from VMS data) within the North Minch (FU11) relative to the areas of the Nature Conservation MPA (NCMPA) and Special Area of Conservation (SAC) showing areas within these where demersal trawling is banned (hatched) and where it is permitted for vessels below 500 kW (clear; depending on gear type, see SG, 2016). Geographic Coordinate System: OSGB 1936, Datum: OSGB 1936, Projected Coordinate System: British National Grid. Coastline by Wessel and Smith (2016), MPA sites subsetted from NCMPA (SNH, 2015) and SAC (SNH, 2016) layers, management areas by SG (2017b) and functional units generated from merged ICES rectangles (ICES, 2017). Map and modified layers created using ArcGIS (ESRI, 2014).

# 13 Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 12 (West of Scotland, South Minch) 

## Type of assessment in 2019

The assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follow the process defined by the benchmark WG (WKNEPH, 2009; WKNEPH, 2013). Full details are provided in the stock annex.

## ICES advice applicable to 2018

'ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2014-2016, catches in 2018 should be no more than 4112 tonnes.

To ensure that the stock in functional unit (FU) 12 is exploited sustainably, management should be implemented at the functional unit level.'

## ICES advice applicable to 2019

'ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2015-2017, catches in 2019 should be no more than 5844 tonnes.

To ensure that the stock in Functional Unit 12 is exploited sustainably, management should be implemented at the functional unit level.'

### 13.1 General

## Stock description

The South Minch (FU12) is located midway down the west coast of Scotland (North Minch report, Section 15, Figure 15.1). The area is characterised by numerous islands of varying size, and sea lochs occur along the mainland coast. These topographical features create a diverse habitat with complex hydrography and a patchy distribution of soft sediments. Further details are provided in the stock annex.

## Management applicable to 2018 and 2019

Management is at the ICES subarea level as described at the beginning of Section 15 (FU11 North Minch report).

## Ecosystem aspects

Details of the ecosystem aspects for this functional unit are provided in the stock annex where available.

## Fishery description

Information on developments in the fishery was provided by Marine Scotland compliance officers. In 2018, the fishery was described as very poor, with many vessels moving over to fish on the east coast of Scotland. Vessels unwilling or unable to move fishing areas are struggling financially.

Two distinct fleets operate in the South Minch and the main ports are Oban and Mallaig. In Oban, there are five local prawn trawlers and ten creelers while there are nine Nephrops trawlers in Mallaig. There were no major changes to the fleets in 2018, but there was a comment that more creels were on the grounds and that the price for brown crab was very good.

Since 2009, vessels have been required to fit 120 mm square meshed panels, in accordance with the west coast emergency measures (Council Reg. (EU) 43/2009). Large SMPs ( 200 mm ) are also widely used and were made mandatory for all TR2 vessels with power $>112 \mathrm{~kW}$ fishing as part of the previous Scottish Conservation Credits scheme. Twin rig vessels tend to use a 200 mm square mesh panel with a 100 mm or larger mesh codend. These vessels do not catchbulk quantities and this leads to Nephrops of better average size and quality. A comment was noted in 2017 about the use of bungee cords to keep the meshes closed. This was investigated by Compliance officers but was deemed to be legal and was not reported as a problem in 2018.

There is very little fish bycatch landed due to the restrictions on cod, haddock and whiting (detailed in ICES, 2016a, ICES, 2016b and ICES, 2016c). Estimates of discard rates of haddock and whiting remain high (ICES, 2016d and ICES, 2017a). Haddock in areas 6 .a are now covered by the landings obligation.
In 2018, there was a large reduction in landings and effort (Figure 16.2.1) in all three functional units on the west coast. This reduction was partly explained by the migration of the west coast fleet to the east coast to take advantage of improved Nephrops fishing opportunities in 4 a . Anecdotal information from the fishing industry suggests that an additional factor contributing to the migration of the fishing fleet was an issue with foreign crew being unable to work in the inshore grounds of the west coast therefore moved to the offshore grounds of the east coast. Further general information on the fishery can be found in the stock annex.

### 13.2 Data available

## InterCatch

Data for 2018 were successfully uploaded into InterCatch prior to the 2019 WG meeting. Uploaded data were worked up in InterCatch to generate 2018 raised international length-frequency distributions. Allocation schemes for any unsampled fleets are described in the stock annex. Data exploration in InterCatch has previously shown that outputs of raised data were very close to those generated by the previous method applied internally with differences being $<0.1 \%$. As such, InterCatch length-frequency outputs have been used in the stock assessment since 2012.

## Commercial catch

Official catch statistics (landings) reported to ICES are shown in Table 15.1.1 (see FU11 North Minch report, Section 15). These relate to the whole of $6 . a$ of which the South Minch is a part. Landings for FU12 provided through national laboratories are presented in Table 16.2.1, broken down by country and by gear type. Landings from this fishery are predominantly reported from

Scotland, with low levels reported from the rest of the UK and Ireland. Total reported Scottish landings in 2018 were 2457 tonnes (plus 45 tonnes from other UK vessels and 34 tonnes from Ireland), consisting of 1692 tonnes ( $77 \%$ ) landed by Scottish Nephrops trawlers and 679 tonnes $(28 \%)$ landed by Scottish creel vessels. The proportion of creel caught landings has remained relatively stable in recent years at $\sim 20 \%$ however, there was an increase in the percentage of creel landings in 2018 to $28 \%$.

## Effort data

In 2015, WGCSE agreed that effort should be reported in kW days as this is likely to be more informative about changes in the actual fleet effort. Effort shows an overall decreasing trend since 2003 but there are peaks in 2008 and 2012, which can be attributed to visiting North Sea trawlers (Figure 16.2.1). There was a decline in effort in 2017 and 2018, attributed to the poor fishing and the movement of vessels to the East coast of Scotland. Note that the effort time-series range (2000-2018) does not match with the more extensive year range available for landings due to a lack of reliable effort data in the Marine Scotland Science in-house database. The effort is also slightly inconsistent with the landings data in that effort is provided for TR2 vessels while the 'Nephrops trawl' landings additionally includes landings by large mesh trawlers targeting Nephrops.

## Sampling levels

Length compositions of landings and discards are obtained during market sampling and onboard observer sampling respectively. These samplinglevels are shown in Table 15.2.2 (seeFU11 North Minch report, Section 15). Sampling effort decreased in 2018 compared to 2017 because of a change in the sampling design. This change was a trail and has since been reverted to the previous design due to the negative effect on sample numbers. This change had no effect on mean weights in landings and discards, these figures fell well within normal ranges. Length compositions for the creel fishery are available for landings only as the small numbers of discardssurvive well. Therefore these animals are not considered to be removed from the population and hence a value of $100 \%$ survival is used (ICES, 2013).

## Length compositions

Figure 16.2.2 shows a series of annual length-frequency distributions from 2000 onwards and appears fairly stable over the time-series. Catch (removals) length compositions are shown for each sex along with the mean size for both. The mean size of males increased slightly in 2018 compared to 2017 in contrast to the females, which showed a decrease. There is little evidence of any recruitment in 2018. The tails of the distributions above 35 mm CL (the size beyond which the effects of recruitment pulses and discards are considered to be negligible) were stable in 2018. It is unclear why there has been an increase in the mean size of the males and a decrease in the mean size of females.

## Sex ratio

The sex ratio in the South Minch shows some variation but males consistently make the largest contribution to the annual landings. Males are available throughout the year while females are mainly caught in the summer when they emerge from the burrow after egg hatching. In 2018, landings were lower in quarter one and four (Figure 16.2.3 (a)). Figure 16.2.3 (b) illustrates the sex ratio by season. There are no particularly anomalous values evident in 2018, although there
was a higher proportion of males in quarter one and a reduction in the male proportion in quarter two of 2018.

## Mean weights

The mean weight in the landings (Figures 15.2.4 and 15.2.5; see FU11 North Minch report, Table 15.2.3) has fluctuated at a high level (in comparison to values for 2006 to 2010) since 2011. Seasonal variability (and occasional outliers) in mean weights is seen in the individual sample estimates. There appears to be a small increase in the mean weight of the males for the trawl caught Nephrops and also, for the females caught by creels (Figure 15.2.5). The annual estimate of mean weight in the landings has an effect on the catch forecast. Over the time-series, there is a general increasing trend in mean weights in the landings. This can be explained by the increasing proportion of creel samples (which tend to catch and land larger Nephrops).

## Discarding

Discarding of undersized and unwanted Nephrops occurs in this fishery. Discard sampling has been conducted on the Scottish Nephropstrawler fleet since 1990. Discarding rates in this FU have varied considerably over the years, ranging from as low as $3 \%$ to over $25 \%$. In 2018, the discarding rate was $4.5 \%$, lower than in 2017 ( $9.1 \%$ ) and is the second lowest discard rate in the timeseries and may be explained because of the poor fishing (Table 16.2.2).

Studies (Charuau et al., 1982; Sangster et al., 1997; Wileman et al., 1999) suggest that some Nephrops survive the discarding process, an estimate of $25 \%$ survival is assumed for this FU in order to calculate removals (landings + dead discards) from the population. The discard survival rate for creel caught Nephrops has been shown to be high (ICES, 2013) and a value of $100 \%$ is used. The discard rate (adjusted to account for survival) which will be used in the forecast was estimated by taking a three-year average 2016-2018 and amounts to 7.3\%.

## Abundance indices from UWTV surveys

Underwater TV surveys using a stratified random approach are available for this stock since 1995. TV surveys are targeted at known areas of mud, sandy mud and muddy sand in which Nephrops construct burrows. The numbers of valid stations used in the final analysis in each year are shown in Table 16.2.3. On average, 35 stations have been considered valid each year, and raised to a stock area of $5072 \mathrm{~km}^{2}$ (derived from BGS sediment data). In 2019, 40 valid stations were used in the survey final analysis (Table 16.2.3).

TV survey abundance estimates from 1999-2019 are shown in Table 16.2.3 and Figure 16.2.4. They show that the Nephrops population in the South Minch experienced several years of high abundance in the early mid-2000s. Aside from this, it has fluctuated without obvious trend over the period of the survey (Figure 16.2.4). The recently observed 2019 abundance represents a $21.37 \%$ increase in relation to 2018.

Table 16.2.4 shows a more detailed summary of the results from the three most recent TV surveys conducted in FU12. The table includes estimates of abundance and variability in each of the strata adopted in the stratified random approach. Mean burrow density increased in 2019, in comparison to the 2018 survey. Densities are generally lower in the western parts of the area towards the Outer Hebrides and higher in the inshore areas to the south west of Skye (Figure 16.2.5). CVs for the three most recent TV surveys (Table 16.2.4) are lower than the precision level agreed by WGNEPS (2019;12.1\%). Figure 16.2 .4 show the time-series estimated abundance for the TV surveys, with $95 \%$ confidence intervals on annual estimates.

The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009; ICES, 2013), WGNEPS (ICES, 2018a), WKNEPS (ICES, 2018b) and (Leocádio et al., 2018). A number of potential biases were highlighted including those due to edge effects, species burrow misidentification and burrow occupancy. The cumulative relative to absolute conversion factor estimated for FU12 was 1.32 meaning that the TV survey is likely to overestimate Nephrops abundance by $32 \%$.

### 13.3 Assessment

## Comparison with previous assessments

The assessment follows the same procedure as last year and is based on a combination of examining trends in fishery indicators and underwater TV abundance estimates. The process was defined by the benchmark WG and is described in the stock annex.

No major issues were highlighted by the audit conducted last year.

## State of the stock

The underwater TV survey is presented as the best available information on the South Minch (FU12) Nephrops stock. The details of the 2019 survey are shown in Table 16.2.4, and compared with the 2017 and 2018 outcomes. At present, it is not possible to extract any length or age structure information from the survey and therefore it provides information on abundance over the area of the survey.

TV survey estimated stock abundance in 2019 was 2362 million individuals, above the MSYB $B_{\text {trigger }}$ value of 1016 million, or the rounded value of 1020 million used for the provision of advice.

The calculated harvest rate in 2019 (dead removals/TV abundance $=4.8 \%$ ) was below the Fmsy proxy for this stock (the value associated with high long-term yield and low risk depletion) of 11.7\%.

### 13.4 Catch scenarios table

Landings predictions and catch options at various harvest rates (based on principles established at WKNEPH (ICES, 2009), will be made on the basis of the 2019 UWTV survey conducted in June. These will be presented in October 2019 for the provision of advice.

Catch scenarios table inputs and historical estimates of mean weight in landings and harvest rates are presented in Table 16.2.2 and summarised below. The calculation of catch scenarios for the South Minch follows the procedure outlined in the stock annex.

Given the variability in mean weights it was considered more appropriate to use a full timeseries average, from 1999 (first year with creel and trawl length distributions combined) until 2018.

The table below shows the agreed inputs to the catch scenario table.

| Input | Data | 2019 assessment |
| :--- | :--- | :---: |
| Survey abundance (millions) | UWTV 2019 | 2362 |
| Mean weight in wanted catch (g) | $1999-2018$ | 26.79 |
| Mean weight in unwanted catch (g) | $1999-2018$ | 10.08 |
| Dead unwanted catch | $2016-2018$ | $7.3 \%$ |
| Discard survival rate | $2016-2018$ | $25 \%$ |

### 13.5 Reference points

New reference points were derived for this stock at WKMSYRef4 (ICES, 2016e). These are updated on the basis of an average of estimated Fmsy proxy harvest rates over a period of years which corresponds more closely to the methodology for finfish. In cases where there is a clear trend in the values, a five-year average was chosen. Similarly, the five-year average of the F at $95 \%$ of the YPR obtained at the FMSY proxy reference point was proposed as the FMSY lower bound and the five-year average of the F above Fmax that leads to YPR of $95 \%$ of the maximum as the upper bound. Using an average value also has the advantage of reducing the effect of any unusually high or low estimates of the Fmsy proxy, which occasionally appear. For this stock, the Fmsy proxy has been revised from $12.3 \%$ to $11.7 \%$.

For Nephrops stocks, MSY Btrigger has been defined as the lowest stock size from which the abundance has increased and is calculated as 1016 million individuals (in 2010). This value was rounded to 1020 million, in the advice from WKMSYRef4 on MSY Btrigger. Full details are contained in the stock annex.

These should remain under review by WGCSE and may be revised should improved data become available.

Table 16.2.2 and Figure 16.5.1 show the harvest rates for FU12. The harvest rate has fluctuated over the time-series and has been below the Fmsy proxy since 2013. The increase in 2016, compared to the 2013-2015 harvest rates, was due to relatively increased landings compared to abundance. The harvest rate has more than halved in 2018 compared to 2017.

It is likely that prior to 2006, the harvestrates are underestimates due to under-reported landings.

### 13.6 Management strategies

Scotland has established a network of regional Inshore Fisheries Groups (rIFGs), non-statutory bodies that aim to improve the management of Scotland's inshore fisheries out to six nautical miles, and to give commercial inshore fishermen a strong voice in wider marine management developments. The rIFGs will contribute to regional policies and initiatives relating to management and conservation of inshore fisheries, including impacts on the marine environment and the maintenance of sustainable fishing communities and measures designed to better conserve and sustainably exploit stocks of shellfish and sea fish (including salmon) in their local waters. Although no IFG proposals specific to the management of Nephrops fisheries have yet been adopted, some of the IFG management plans for the Scottish West Coast include spatial management of Nephrops fisheries and the introduction of creel limits.

On the 8th of February 2016, phase 1 of the fisheries management measures for inshore MPAs in Scottish waters came into force (SG, 2016). These measures relate to both Nature Conservation

MPAs (NCMPAs; Marine (Scotland) Act and the UK Marine and Coastal Access Act) and Special Areas of Conservation (SACs; EC Habitats Directives - Council Directive 92/43/EEC) both of which have the aim of conserving biological diversity in Scottish waters and contribute to Scotland's MPA network (SG, 2017a). Although not specific to the management of the Nephrops fishery, they will influence spatial patterns of fishing for Nephrops where controls on the two main gear types, demersal trawls and creels, are implemented on Nephrops habitat. There are seven protected areas within the South Minch functional unit with fisheries management measures. MPAs on the main areas of Nephrops habitat include the Loch Sunart to the Sound of Jura NCMPA where demersal trawling is banned in some areas, i.e. zoned, and seasonal closures implemented in others, Loch Sunart NCMPA/SAC, where demersal trawling is banned and creeling is zoned, the East of Mingulay SAC, demersal trawling banned and creeling zoned, and the Trenish Isles SAC, demersal trawlingbanned. Another area is the LochDuich, Long and Alsh NCMPA/SAC, covering some patches of muddy sediment, where demersal trawling is banned or temporally closed in other areas that extend beyond the MPA onto muddy sediment. Other areas include the Loch CreranSAC/NCMPA, demersal trawling banned and creeling zoned, and the Firth of Lorn SAC, which has the same management as the Loch Sunart to the Sound of Jura NCMPA. For the Firth of Lorn and Loch Creran, management was in place prior to 2016 (SG, 2016). An additional NCMPA, at Loch Carron, was designated using emergence powers in 2017 (SG, 2017b). The areas of the SACs and NCMPAs relative to the estimated Nephrops habitat within the South Minch functional unit are displayed in Figure 16.6.1.

Also, need to add EU MAP WW 2019 in Section on Nephrops 11 as management details and have it referenced also.

### 13.7 Quality of assessment and forecast

The length and sex composition of the landings data is considered to be adequately sampled, although sampling levels were lower in 2018 compared to 2017 (see Section 16.2). Discard sampling has been conducted for Scottish Nephrops trawlers in this fishery since 1990, and is considered to represent the trawl fishery adequately. The landings length compositions from 1999 onwards are derived from both creel and trawl samples. The creel fishery, which generally accounts for around $20 \%$ of the landings (although increased to $27 \%$ in 2018) and increasingly operates over similar areas to trawling, exhibits a length composition composed of larger animals.

There are concerns over the accuracy of historical landings and effort data prior to 2006 when Buyers and Sellers legislation was introduced and the reliability began to improve. Because of this, the final assessmentadopted is independent of official statistics. Harvest rates since 2006 are also considered more reliable due to more accurate landings data reported under new legislation. Incorporation of creel length compositions has also improved estimates of harvest rates.

Underwater TV surveys have been conducted for this stock every year since 1995. The number of valid stations in the survey has remained relatively stable throughout the time period. The survey is targeted at known areas of mud, sandy mud and muddy sand within the South Minch. The variance of density estimates in the South Minch is relatively high, particularly in the sandy mud strata, resulting in large confidence intervals and a greater uncertainty on the abundance estimates than in other FUs. This makes it difficult to determine which population changes are significant. Although the CV's have been smaller in recent years.

There is a need to explore options to implement further stratification for the South Minch survey area. In the provision of catch options based on the absolute survey estimates, additional uncertainties related to mean weight in the landings and the discard rates also arise. A three-year average (2016-2018) of discard rates (adjusted to account for some survival of discarded animals) has been used in the calculation of catch options.

The cumulative relative to absolute conversion factor estimates for FU12 are largely based on expert opinion. The precision of these bias corrections cannot yet be characterised. The landings derived in the forecast (catch options table) are sensitive to the input dead discard rate and mean weights in landings, and this introduces uncertainties in the catch forecasts. Precision estimates are needed for these forecast inputs.

The overall area of the ground is estimated from the available BGS contoured sediment data and at present is considered to be a minimum estimate. Work is underway to improve the area estimation. VMS data linked to landings (from queries of the Scottish FIN database) suggest no major differences between areas fished and the mud sediment maps. Two other factors however, are likely to increase the estimate of ground area available for Nephrops and Nephrops directed fishing. Firstly, the inclusion of vessels smaller than 15 m would likely increase the fished area in some of the inshore locations and secondly, it is known that most of the sea lochs have areas of mud substrate and are typically fished by creel boats. In recent years, limited TV surveys have taken place in some of the sea lochs and attempts are being made to utilise these data to improve estimates of mud area and Nephrops abundance in the South Minch.

### 13.8 Recommendation for next benchmark

This stock was last benchmarked in 2009. WGCSE will keep the stock under close review and recommend future benchmark as required.

### 13.9 Management considerations

ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES division level. Management at the functional unit level could provide controls to ensure effort and catch were in line with resources available.

Creel fishing takes place in this area but overall effort in terms of creel numbers is not known and measures to control numbers are not in place. There is a need to ensure that the combined effort from all forms of fishing is taken into account when managing this stock.

There is a bycatch of other species in the area of the South Minch and estimated discards of whiting and haddock by the TR2 fleet are high in area 6.a generally. It is important that efforts continue to ensure that unwanted bycatch is kept to a minimum in this fishery.

### 13.10 References

Afonso-Dias, M. S. 1998. Variability of Nephrops norvegicus (L.) populations in Scottish waters in relation to the sediment characteristics of the seabed. PhD Thesis University of Aberdeen. 282 pp.

BGS. 2011. Marine SeaBed Sediment Map - UK Waters - 250k (DigSBS250). British Geological Survey, Nottingham. Available at: [http://www.bgs.ac.uk/discoverymetadata/13605549.html](http://www.bgs.ac.uk/discoverymetadata/13605549.html) [Accessed: unknown date].

Charuau A., Morizur Y., Rivoalen J.J. 1982. Survival of discarded Nephrops norvegicus in the Bay of Biscay and in the Celtic Sea, ICES-CM-1982/B:13.

Council Reg. (EU) 43/2009.
EU. 2015. COMMISSION DELEGATED REGULATION (EU) 2015/2438 of 12 October 2015 establishing a discard plan for certain demersal fisheries in north-western waters. Official Journal of the European Union, L 336/29. Available at: [http://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32015R2438\&from=EN](http://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32015R2438%5C&from=EN) [Accessed: 30/06/2017].

ESRI. 2014. ArcGIS. Version 10.2.1. Environmental Systems Research Institute, Inc.: Redlands, CA.

Howard F.G. and Hall, W.B. 1983. Some observations on the biometrics of Nephrops norvegicus (L.) in Scottish waters. ICES, Doc.ShellfishComm.,CM1983/K:36.

ICES. 2010. Report of the Study Group on Nephrops Surveys (SGNEPS), 9-11 November 2010, Lisbon, Portugal. ICES CM 2010/SSGESST:22. 95 pp.
ICES. 2012. Report of the Study Group on Nephrops Surveys (SGNEPS), 6-8 March 2012, Acona, Italy. ICES CM 2012/SSGESST:19. 36 pp.

ICES. 2013. Report of the Benchmark Workshop on Nephrops assessment (WKNEPH). ICES CM 2013/45 230 pp .
ICES. 2016a. Stock Annex: Whiting (Merlangius merlangus) in Division 6.a (West of Scotland). Working Group for the Celtic Seas Ecoregion (WGSCE). Available at: [http://ices.dk/sites/pub/Publica-tion\ Reports/Stock\ Annexes/2016/whg-scow_SA.pdf.](http://ices.dk/sites/pub/Publica-tion%5C%20Reports/Stock%5C%20Annexes/2016/whg-scow_SA.pdf.) [Accessed: 30/06/2017].

ICES. 2016b. Stock Annex: West of Scotland cod (Division 6.a). Working Group for the Celtic Seas Ecoregion (WGSCE). Available at: [http://ices.dk/sites/pub/Publication\ Reports/Stock\ Annexes/2016/codscow_SA.pdf.](http://ices.dk/sites/pub/Publication%5C%20Reports/Stock%5C%20Annexes/2016/codscow_SA.pdf.) [Accessed: 30/06/2017].
ICES. 2016c. Haddock (Melanogrammusaeglefinus) in Subarea 4, Division 6.a and Subdivision 20 (North Sea, West of Scotland, Skagerrak). Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK). Available at: [http://www.ices.dk/sites/pub/Publication\ Re-ports/Stock\ Annexes/2016/had-346a_SA.pdf.](http://www.ices.dk/sites/pub/Publication%5C%20Re-ports/Stock%5C%20Annexes/2016/had-346a_SA.pdf.) [Accessed: 30/06/2017].

ICES. 2016d. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 4-13 May 2016, Copenhagen, Denmark. ICES CM 2016/ACOM:13. 1343 pp.
ICES.2016e. EU request to ICES to provide FMSY ranges for selected stocks in ICES subareas 5 to 10. In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.2.3.1.

ICES. 2017a. Report of the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), 26 April-5 May 2016, Hamburg, Germany. ICES CM 2016/ ACOM:14. 19 pp.

ICES. 2017b. ICES Statistical Rectangles. International Council for the Exploration of the Sea: Copenhagen, Denmark. Available at: [http://geo.ices.dk/index.phpSource](http://geo.ices.dk/index.phpSource) [Accessed: 16/03/2017].
ICES. 2018a. Report of the Working Group on Nephrops Surveys (WGNEPS), 6-8 November 2018. Lorient, France. ICES CM 2018/ EOSG:18.227 pp.
ICES. 2018b. Report of the Workshop on Nephrops burrow counting, 2-5 October 2018, Aberdeen, UK. ICES CM 2018/ EOSG:25.47 pp.
Leocádio, A., Weetman, A., and Wieland, K. (Eds). 2018. Using UWTV surveys to assess and advise on Nephrops stocks. ICES Cooperative Research Report No. 340. 49 pp. http://doi.org/10.17895/ices.pub. 4370.

Sangster, G.I., Breen, M., Bova, D.J., Kynoch, R., O’Neill, F.G., Lowry. N., Moth-Poulsen, T. Hansen, U.J., Ulmestrand, M., Valentinsson, D., Hallback, H., Soldal, A.V., and Hoddevik, B. 1997. Nephrops survival after escape and discard from commercial fishing gear. Presented at ICES FTFB Working Group, Hamburg, Germany 14-17 April, 1997, ICES CM 1997 CM/B.

SG. 2016. Simpleguide to fisheries management measures in Marine Protected Areas. Marine Scotland (The Scottish Government): Edinburgh. Available at: [http://www.gov.scot/Resource/0049/00498320.pdf](http://www.gov.scot/Resource/0049/00498320.pdf) [Accessed: 16/05/2017].
SG. 2017a. Marine Protected Areas in Scotland's Seas - Guidelines on the selection of MPAs and development of the MPA network. Marine Scotland (The Scottish Government): Edinburgh. Available at: [http://www.gov.scot/Topics/marine/marine-environment/mpanetwork/mpaguidelines](http://www.gov.scot/Topics/marine/marine-environment/mpanetwork/mpaguidelines) [Accessed: 16/05/2017].

SG. 2017b. Loch Carron Marine Protected Area Question and Answer Document. Marine Scotland (The Scottish Government): Edinburgh. Available at: [http://www.gov.scot/Resource/0051/00518275.pdf](http://www.gov.scot/Resource/0051/00518275.pdf) [Accessed: 01/06/2017].

SG. 2017c. Marine conservation orders (MCOs) and fisheries management measures (MPAs and SACs) with effect May 2017. Marine Scotland (The Scottish Government): Edinburgh. Available at: [https://marinescotland.atkinsgeospatial.com/nmpi/default.aspx?availablelayers=838](https://marinescotland.atkinsgeospatial.com/nmpi/default.aspx?availablelayers=838) [Accessed: 31/05/2017].

SNH. 2015. Nature Conservation Marine Protected Areas. Scottish Natural Heritage (Scottish Government): Inverness. Available at: [https://gateway.snh.gov.uk/natural-spaces/index.jsp](https://gateway.snh.gov.uk/natural-spaces/index.jsp) [Accessed: 03/04/2017].

SNH. 2016. Special Areas of Conservation. Scottish Natural Heritage (Scottish Government): Inverness. Available at: [https://gateway.snh.gov.uk/natural-spaces/index.jsp](https://gateway.snh.gov.uk/natural-spaces/index.jsp) [Accessed: 03/04/2017].
Wessel, P. and Smith, W.H.F. 2016. GSHHG version 2.3.6-A Global Self-consistent, Hierarchical, Highresolution Geography Database. National Centres for Environmental Information, National Oceanic and Atmospheric Administration: Boulder, CO. [https://www.ngdc.noaa.gov/mgg/shorelines/gshhs.html](https://www.ngdc.noaa.gov/mgg/shorelines/gshhs.html) [Accessed: 31/05/2017].

Wileman, D.A., Sangster, G.I., Breen, M., Ulmestrand, M., Soldal, A.V. and Harris, R.R. 1999. Roundfish and Nephrops survival after escape from commercial fishing gear. EU Contract Final Report. EC Contract No: FAIR-CT95-0753.

Table 16.2.1. Nephrops, South Minch (FU12), ICES estimates of landings of Nephrops, 1981-2018.

| UK SCOTLAND |  |  |  |  |  | OTHER UK | IRELAND | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | NEPHROPS TRAWL | OTHER | CREEL | BELOW MINIUM SIZE | SUB TOTAL |  |  |  |
| 1981 | 2966 | 254 | 432 | 0 | 3652 | 0 | 0 | 3652 |
| 1982 | 2925 | 206 | 421 | 0 | 3552 | 0 | 0 | 3552 |
| 1983 | 2595 | 362 | 456 | 0 | 3413 | 0 | 0 | 3413 |
| 1984 | 3229 | 477 | 594 | 0 | 4300 | 0 | 0 | 4300 |
| 1985 | 3096 | 424 | 488 | 0 | 4008 | 0 | 0 | 4008 |
| 1986 | 2694 | 288 | 502 | 0 | 3484 | 0 | 0 | 3484 |
| 1987 | 2928 | 418 | 546 | 0 | 3892 | 0 | 0 | 3892 |
| 1988 | 3544 | 364 | 555 | 0 | 4463 | 10 | 0 | 4473 |
| 1989 | 3846 | 338 | 561 | 0 | 4745 | 0 | 0 | 4745 |
| 1990 | 3732 | 263 | 435 | 0 | 4430 | 0 | 0 | 4430 |
| 1991 | 3596 | 342 | 503 | 0 | 4441 | 1 | 0 | 4442 |
| 1992 | 3478 | 209 | 549 | 0 | 4236 | 1 | 0 | 4237 |
| 1993 | 3609 | 194 | 650 | 0 | 4453 | 5 | 0 | 4458 |
| 1994 | 3742 | 264 | 405 | 0 | 4411 | 3 | 0 | 4414 |
| 1995 | 3443 | 717 | 508 | 0 | 4668 | 14 | 0 | 4682 |
| 1996 | 3108 | 417 | 469 | 0 | 3994 | 1 | 0 | 3995 |
| 1997 | 3518 | 329 | 493 | 0 | 4340 | 3 | 1 | 4344 |
| 1998 | 2851 | 340 | 538 | 0 | 3729 | 0 | 1 | 3730 |
| 1999 | 3165 | 359 | 514 | 0 | 4038 | 0 | 14 | 4052 |
| 2000 | 2940 | 311 | 700 | 0 | 3951 | 0 | 2 | 3953 |
| 2001 | 2823 | 391 | 768 | 0 | 3982 | 0 | 9 | 3991 |
| 2002 | 2234 | 314 | 743 | 0 | 3291 | 0 | 14 | 3305 |
| 2003 | 2812 | 203 | 858 | 0 | 3873 | 0 | 6 | 3879 |
| 2004 | 2864 | 105 | 879 | 0 | 3848 | 0 | 21 | 3869 |
| 2005 | 2812 | 46 | 955 | 0 | 3813 | 1 | 34 | 3848 |
| 2006 | 3570 | 97 | 922 | 0 | 4589 | 9 | 35 | 4633 |
| 2007 | 4437 | 21 | 959 | 0 | 5417 | 19 | 35 | 5471 |


| UK SCOTLAND |  |  |  |  |  | OTHER UK | IRELAND | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | NEPHROPS TRAWL | OTHER | CREEL | BELOW MINIUM SIZE | SUB TOTAL |  |  |  |
| 2008 | 4433 | 12 | 896 | 0 | 5341 | 2 | 13 | 5356 |
| 2009 | 3346 | 24 | 900 | 0 | 4270 | 4 | 11 | 4285 |
| 2010 | 2836 | 19 | 969 | 0 | 3824 | 16 | 6 | 3846 |
| 2011 | 2876 | 11 | 783 | 0 | 3670 | 23 | 9 | 3702 |
| 2012 | 3159 | 32 | 773 | 0 | 3964 | 19 | 6 | 3989 |
| 2013 | 2490 | 543 | 729 | 0 | 3762 | 13 | 1 | 3776 |
| 2014 | 2490 | 3 | 637 | 0 | 3130 | 32 | 17 | 3179 |
| 2015 | 2662 | 18 | 665 | 0 | 3345 | 22 | 33 | 3400 |
| 2016 | 3450 | 22 | 838 | 0 | 4310 | 33 | 59 | 4402 |
| 2017 | 2741 | 54 | 768 | 0 | 3563 | 23 | 66 | 3652 |
| 2018 | 1692 | 86 | 679 | 0 | 2457 | 45 | 34 | 2536 |

Table 16.2.2. Nephrops, South Minch (FU12): Adjusted TV survey abundance, landings, discard rate proportion by number) and estimated harvest rate.

| YEAR LAND- <br>  INGS <br>  NUM- <br>  BER <br>  (MIL- <br>  LIONS) | DIS- <br> CARDS <br> NUM- <br> BER <br> (MIL- <br> LIONS) |  | REMOV- <br> ALS <br> NUM- <br> BER <br> (MIL- <br> LIONS)* | ADJUSTED SURVEY (MILLIONS) | HARVEST RATE* | LANDINGS (TONNE S) | DISCARDS (TONNE S) | DISCARD RATE <br> (\%) | DEAD <br> DISCARD <br> RATE <br> (\%) | MEAN <br> D WEIGHT <br> IN <br> LAND- <br> INGS (g) | MEAN WEIGHT IN DISCARDS (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 161 | 29 | 183 | 1086 | 16.9 | 4052 | 206 | 15.4 | $12 \quad 2$ | 25.14 | 7 |
| 2000 | 145 | 33 | 170 | 1854 | 9.2 | 3953 | 284 | 18.7 | 14.7 | 27.3 | 8.5 |
| 2001 | 168 | 65 | 216 | 2037 | 10.6 | 3991 | 591 | 27.9 | 22.5 | 23.79 | 9.11 |
| 2002 | 123 | 26 | 143 | 1899 | 7.5 | 3305 | 247 | 17.6 | 13.8 | 26.83 | 9.37 |
| 2003 | 139 | 38 | 168 | 2157 | 7.8 | 3879 | 381 | 21.3 | 16.9 | 27.86 | 10.1 |
| 2004 | 141 | 44 | 175 | 2558 | 6.8 | 3869 | 454 | 23.8 | 19 | 27.37 | 10.26 |
| 2005 | 137 | 49 | 174 | 2208 | 7.9 | 3848 | 452 | 26.5 | 21.2 | 28.11 | 9.17 |
| 2006 | 177 | 30 | 199 | 1845 | 10.8 | 4633 | 324 | 14.3 | 11.1 | 26.24 | 10.97 |
| 2007 | 228 | 66 | 278 | 1016 | 27.3 | 5471 | 903 | 22.4 | 17.8 | 23.95 | 13.73 |
| 2008 | 224 | 74 | 279 | 1608 | 17.4 | 5356 | 605 | 24.7 | 19.8 | 23.91 | 8.23 |
| 2009 | 179 | 26 | 199 | 1542 | 12.9 | 4285 | 216 | 12.5 | 9.6 | 23.87 | 8.44 |
| 2010 | 149 | 12 | 158 | 2076 | 7.6 | 3846 | 133 | 7.7 | 5.9 | 25.86 | 10.76 |
| 2011 | 118 | 11 | 126 | 1945 | 6.5 | 3702 | 92 | 8.2 | 6.3 | 31.1 | 8.78 |
| 2012 | 133 | 16 | 145 | 919 | 15.8 | 3989 | 145 | 10.8 | 8.3 | 29.17 | 9.05 |
| 2013 | 136 | 4 | 140 | 1718 | 8.1 | 3776 | 50 | 3.1 | 2.4 | 27.48 | 11.31 |
| 2014 | 105 | 19 | 120 | 2073 | 5.8 | 3179 | 233 | 15.6 | 12.1 | 29.91 | 12.04 |
| 2015 | 120 | 10 | 128 | 1998 | 6.4 | 3400 | 121 | 7.7 | $5.9 \quad 2$ | 28.15 | 12.04 |
| 2016 | 177 | 31 | 201 | 2118 | 9.5 | 4402 | 365 | 14.9 | 11.6 | 24.76 | 11.74 |
| 2017 | 127 | 13 | 137 | 1384 | 9.9 | 3652 | 105 | 9.1 | 7 | 27.76 | 8.29 |
| 2018 | 91 | 4 | 94 | 1946 | 4.8 | 2536 | 54 | 4.5 | $3.4 \quad 27$ | 27.27 | 12.74 |
| Average*** |  |  |  |  |  |  |  |  | 7.3 | 26.79 | 10.08 |

*Harvest rates previous to 2006 are unreliable.
**Removals numbers take the dead discard rate into account.
*** Dead discard average: 2016-2018; Mean weight in landings and discards average: 1999-2018.

Table 16.2.3. Nephrops, South Minch (FU12): Results of the 1995-2019 TV surveys (adjusted for bias).

| YEAR | NUMBER OF VALID STATIONS | MEAN DENSITY (BURROWS/m ${ }^{2}$ ) | ABUNDANCE (MILLIONS) | 95\% CONFIDENCE INTERVAL (MILLIONS) |
| :---: | :---: | :---: | :---: | :---: |
| 1995 | 33 | 0.227 | 1152 | 251 |
| 1996 | 21 | 0.288 | 1473 | 530 |
| 1997 | 36 | 0.212 | 1086 | 185 |
| 1998 | 38 | 0.288 | 1452 | 232 |
| 1999 | 37 | 0.212 | 1086 | 260 |
| 2000 | 41 | 0.364 | 1854 | 348 |
| 2001 | 47 | 0.402 | 2037 | 459 |
| 2002 | 31 | 0.371 | 1899 | 567 |
| 2003 | 25 | 0.424 | 2157 | 756 |
| 2004 | 38 | 0.508 | 2558 | 473 |
| 2005 | 33 | 0.432 | 2208 | 740 |
| 2006 | 36 | 0.364 | 1845 | 598 |
| 2007 | 39 | 0.197 | 1016 | 155 |
| 2008 | 33 | 0.318 | 1608 | 415 |
| 2009 | 25 | 0.303 | 1542 | 634 |
| 2010 | 34 | 0.409 | 2076 | 665 |
| 2011 | 36 | 0.383 | 1945 | 778 |
| 2012 | 38 | 0.182 | 919 | 185 |
| 2013 | 38 | 0.339 | 1718 | 365 |
| 2014 | 36 | 0.409 | 2073 | 530 |
| 2015 | 35 | 0.394 | 1998 | 514 |
| 2016 | 37 | 0.417 | 2118 | 440 |
| 2017 | 41 | 0.273 | 1384 | 282 |
| 2018 | 39 | 0.383 | 1946 | 371 |
| 2019 | 40 | 0.466 | 2362 | 578 |

Table 16.2.4. Nephrops South Minch (FU12). Results by stratum of the 2017-2019 TV surveys. Note that stratification was based on a series of sediment strata (M - Mud, SM - Sandy mud, MS - Muddy sand).

| STRATUM | $\begin{aligned} & \text { AREA } \\ & \left(\mathbf{k m}^{2}\right) \end{aligned}$ | NUMBER OF STATIONS | MEAN BURROW DENSITY (no./m²) | OBSERVED <br> VARIANCE | ABUN- <br> DANCE <br> (MIL- <br> LIONS) | STRATUM VARIANCE | PROPOR- <br> TION OF TOTAL VARIANCE | SURVEY <br> PRECISION LEVEL <br> (RSE) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 TV <br> Survey |  |  |  |  |  |  |  |  |
| M | 303 | 0.216 | 0.001 | 65.5 | 65 | 0.003 | 0.216 |  |
| SM | 2741 | 0.314 | 0.037 | 861.1 | 14725 | 0.739 | 0.314 |  |
| MS | 2028 | 20 | 0.226 | 0.025 | 457.7 | 5142 | 0.258 |  |
| Total | 5071 | 41 |  |  | 1384.2 | 19932 | 1 | 0.136 |
| $2018 \text { TV }$ <br> Survey |  |  |  |  |  |  |  |  |
| M | 303 | 2 | 0.311 | 0.007 | 94.2 | 2 | 0.311 |  |
| SM | 2741 | 19 | 0.441 | 0.074 | 1207.5 | 19 | 0.441 |  |
| MS | 2028 | 18 | 0.317 | 0.021 | 644.3 | 18 | 0.317 |  |
| Total | 5071 | 39 |  |  | 1946 | 34438 | 1 | 9.1 |
| $2019 \text { TV }$ <br> Suvey |  |  |  |  |  |  |  |  |
| M | 303 | 2 | 0.466 | 0.001 | 141.2 | 65 | 0.001 |  |
| SM | 2741 | 20 | 0.494 | 0.162 | 1352.7 | 61024 | 0.73 |  |
| MS | 2028 | 18 | 0.428 | 0.099 | 867.9 | 22546 | 0.27 |  |
| Total | 5071 | 40 |  |  | 2361.7 | 83635 |  | 0.121 |

## Landings - International



Effort - Scottish Nephrops trawlers


Figure 16.2.1. Nephrops, South Minch (FU12). Long-term landings and effort.

## Length frequencies for catch (dotted) and landed(solid): Nephrops in FU12



Figure 16.2.2. Nephrops. South Minch (FU12). Catch length-frequency distribution and mean size in catches (dotted) and landings (solid) for Nephrops in the North Minch, 2000-2018. Vertical lines are minimum conservation reference size ( 20 mm ) and 35 mm .


Figure 16.2.3. (a) Nephrops, South Minch (FU12). Landings by sex and quarter from Scottish trawlers.


Figure 16.2.3 (b) Nephrops, South Minch (FU12), Proportion of males by quarter (1980-2018).

## South Minch



Figure 16.2.4. Nephrops, South Minch (FU12), Time-series of TV survey abundance estimate (adjusted for bias), with $95 \%$ confidence intervals, 1995-2019. The dashed blue line is the rounded Btrigger value of $\mathbf{1 0 2 0}$ million individuals.


Figure 16.2.5. Nephrops, South Minch (FU12), TV survey station distribution and relative density (burrows $/ \mathrm{m}^{2}$ ), 2014-2019. Shaded green and brown areas represent areas of suitable sediment for Nephrops. Bubbles in this figure are all scaled the same. Red crosses represent zero observations.

Harvest Rate


Figure 16.5.1. Nephrops, South Minch (FU12), harvest rate, 1995-2019. The dashed and solid lines are the FMSY proxy harvest rate $(\mathbf{1 1 . 7 \%}$ ) and the harvest rate respectively. Harvest rates prior to 2006 are unreliable.


Figure 16.6.1. The area of Nephrops habitat (Mud, Muddy Sand and Sandy Mud) within the South Minch (FU12) relative to the areas of the Nature Conservation MPAs (NCMPAs) and Special Area of Conservations (SACs) with fisheries management measures. Areas where demersal trawling is prohibited, restricted (i.e. vessel size restrictions or seasonal closures) and where creeling is prohibited are displayed. For more detailed information see SG (2016). Geographic Coordinate System: OSGB 1936, Datum: OSGB 1936, Projected Coordinate System: British National Grid. Coastline by Wessel and Smith (2016), MPA sites subsetted from NCMPA (SNH, 2015) and SAC (SNH, 2016) layers, management areas from SG (2017c) and functional units generated from merged ICES rectangles (ICES, 2017b). Map and modified layers created using ArcGIS (ESRI, 2014).

# 14 Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 13 (West of Scotland, the Firth of Clyde and Sound of Jura) 

## Type of assessment in 2019

The assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the benchmark WG (WKNEPH, 2009; WKNEPH, 2013). Full details are provided in the stock annex.

## ICES advice applicable to 2018

'ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2014-2016, catches in 2018 should be no more than 5179 tonnes ( 4484 tonnes for the Firth of Clyde and 695 tonnes for the Sound of Jura).
To ensure that Nephrops stocks are exploited sustainably, management of Nephrops in general should be implemented at the functional unit (FU) level. In this particular FU additional measures may be required to ensure that the landings taken in each subarea (Firth of Clyde and Sound of Jura) are in line with the advice.'

## ICES advice applicable to 2019

'ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2015-2017, catches in 2019 should be no more than 6588 tonnes ( 5990 tonnes for the Firth of Clyde and 598 tonnes for the Sound of Jura).
To ensure that Nephrops stocks are exploited sustainably, management of Nephrops in general should be implemented at the functional unit level. In this particular functional unit additional measures should be implemented to ensure that landings taken in each subarea (Firth of Clyde and Sound of Jura) are in line with the advice.'

### 14.1 General

## Stock description

The Clyde functional unit (FU13) is located in the southern waters off the west coast of Scotland (FU11 report,Section 12,Figure 15.1)). It is comprised of two distinct patches in the Firth of Clyde and the Sound of Jura, to the east and west of the Mull of Kintyre respectively. The hydrography of the two subareas differs, with the Sound of Jura characterised by stronger tidal currents and the Firth of Clyde exhibiting features of a lower energy environment with a shallowentrance sill. Owing to its burrowing behaviour, the distribution of Nephrops is restricted to areas of mud, sandy mud and muddy sand. Within the two distinct patches, these substrates are distributed according to prevailing hydrographic and bathymetric conditions. The available area of suitable sediment is smaller in the Sound of Jura, occupying only the deepest parts of the Sound, while in the Firth of Clyde these sediments predominate. Further details are provided in the stock annex.

## Management applicable to 2018 and 2019

Management is at the ICES subarea level as described at the beginning of Section 12 (FU11 North Minch report).

## Ecosystem aspects

Details of the ecosystem aspects for this functional unit are provided in the stock annex where available.

## Fishery description

Information on developments in the fishery was provided by Marine Scotland fishery compliance officers.

There are 76 resident vessels based in the Campbeltown district, and an even split between trawlers and creel vessels. Vessel power was between $80-585 \mathrm{~kW}$, the most powerful being the over 10 m trawlers. In addition, there are 30 trawlers and 15 creelers of $10-18 \mathrm{~m}$ fishing out of Ayr, with vessel power between $75-300 \mathrm{~kW}$. All trawlers use 80 mm single or twin rigs with square mesh panels (SMP) of at least 120 mm , in accordance with west coast emergency measures conditions (Council Reg. (EU) 43/2009). Under the Scottish Conservation Credits scheme, vessels with power $>112 \mathrm{~kW}$ are required to use a 200 mm SMP.

The activity of Northern Irish vessels was not perceived to be high in 2017 or 2018; vessels did not land locally but went back to their homeport because of better fuel and market prices.

Mobile gear is banned in the Inshore Clyde from Friday night to Sunday night, as are vessels greater than 21 m in length. Most creel boats operating in the Clyde have two crew members and operate around 1000 creels. Creeling activity now takes place quite widely in the northern parts of the Firth operating on some of the same grounds but often taking place during the weekend trawling ban.
In terms of the influence of Marine Protected Area (MPA) management measures on the fishery, it was stated that the South Arran Nature Conservation MPA (NCMPA) removed a large sea area for Nephrops trawlers to operate over. This reportedly increased trawling effort outside of prohibited area. However, this allowed creelers to move into the areas were trawling was banned. There have been recent reports of increases in creel numbers in this area and this has resulted in gear conflict within the creel sector. The small area of the Upper Loch Fyne NCMPA closed to trawlers was reported to have had little impact.

In 2018, there was a large reduction in landings and effort (Figure 17.2.1) in all three functional units on the west coast. This reduction was partly explained by the migration of the west coast fleet to the east coast to take advantage of improved Nephrops fishing opportunities in 4 a . Anecdotal information from the fishing industry suggests that an additional factor contributing to the migration of the fishing fleet was an issue with foreign crew being unable to work in the inshore grounds of the west coast therefore moved to the offshore grounds of the east coast. Further general information on the fishery can be found in the stock annex.

### 14.2 Data available

## InterCatch

Data for 2018 were successfully uploaded into InterCatch prior to the 2019 WG meeting according with the deadline proposed. Uploaded data were worked up in InterCatch to generate 2018 raised international length-frequency distributions. Data exploration in InterCatch has previously shown that outputs of raised data were very close to those generated by the previous method applied internally with differences being $<0.1 \%$. As such, InterCatch length-frequency outputs have been used in the stock assessment since 2012.

## Commercial catch

Official catch statistics (landings) reported to ICES are shown in Table 15.1.1 (see FU11 North Minch report, Section 12). These relate to the whole of area $6 . a$ of which the FU13 is a part. Landings statistics for FU13 provided through national laboratories are presented in Table 17.1.1, broken down by country and by gear type. Landings from this fishery are predominantly reported from Scotland, although Northern Ireland contributed 303 tonnes in 2018. Total reported Scottish landings in 2018 were 3838 tonnes (plus 303 tonnes from other UK vessels i.e. Northern Ireland), consisting of 3584 tonnes landed by trawlers ( $93.3 \%$ ) and 251 tonnes ( $6.5 \%$ ) landed by Scottish creel vessels. Creel landings have generally increased in the most recent years, from approximately $3 \%$ in 2012 to nearly $7 \%$ of total landings in 2018, which is still low compared to the other FUs in the west of Scotland.

Statistical rectangle 40E4 covers parts of both the Firth of Clyde and the Sound of Jura. Table 17.2.1 shows the split in landings between the two subareas comprising FU13. Historically the allocation of landings to the two components of FU13 was carried out by the fishery office and required them to have detailed knowledge of where vessels have been fishing within 40 E 4 . The apparent sudden decline in landings from the Sound of Jura in 2001 is not considered to be associated with a sudden change in fishing practices and is thought more likely to be due to changes in fishery office recording practices. For this reason, the landings split is considered unreliable in recent years and the commercial landings data are now presented for the combined Firth of Clyde and Sound of Jura. Given the relative magnitudes of the fisheries (Clyde likely to be much bigger), the commercial data are likely to be more representative of the Clyde.

## Effort data

In 2015, WGCSE agreed that effort should be reported in kW days, as this is likely to be more informative about changes in the actual fleet effort. Effort shows an overall decreasing trend but was stable through 2010 to 2012 (Figure 17.2.1). Effort increased in 2016 in comparison to 2015 but declined in 2017 and 2018. Note that the effort time-series range (2000-2018) does not match with the more extensive year range available for landings due to a lack of reliable effort data in the Marine Scotland Science in-house database. The effort is also slightly inconsistent with the landings data in that effort is provided for TR2 vessels while the 'Nephrops trawl' landings additionally includes landings by large mesh trawlers targeting Nephrops.

## Sampling levels

Length compositions of landings and discards are obtained during market and on-board observer sampling respectively. These sampling levels are shown in Table 15.2.2 (see FU11 North

Minch report, Section 12). Sampling effort decreased in 2018 compared to 2017 because of a change in the sampling design. This change was a trail and has since been reverted to the previous design due to the negative effect on sample numbers. This change had no effect on mean weights in landings and discards, these figures fell well within normal ranges.

Sampling of landings length compositions in the Sound of Jura is more infrequent but samples have been included in the FU13 raising procedure when available. Length compositions for the creel fishery are available for landings only. This is because survival in the, probably, low numbers of animals that are discarded (although little quantitative information exists) has been shown to be high (ICES, 2013). Therefore these animals are not considered to be removed from the population and hence a value of $100 \%$ survival is used (ICES, 2013).

## Length compositions

Although assessments based on detailed catch analysis are not presently carried out, examination of length compositions can provide a preliminary indication of exploitation effects. Figure 17.2.2 shows a series of annual Clyde length-frequency distributions for the period 2000 to 2018. Catch (removals) length compositions are shown for each sex along with the mean size for both. There has been a decline in the mean sizes of both sexes since 2015. Examination of the tails of the distributions above 35 carapace length CL mm (the length beyond which the effects of recruitment pulses and discards are considered to be negligible) shows the maximum sampled size for both sexes has fallen. However, there is no obvious evidence of over-exploitation of the stock.

## Sex ratio

Sex ratio in the Clyde shows some variation but males generally make the largest contribution to the annual landings (Figure 17.2.3(a)). This occurs because males are available throughout the year and the fishery takes place in all quarters, although effort is generally reduced during the winter months because of poor weather. Females on the other hand are mainly taken in the summer when they emerge after egg hatching. The seasonal change in proportion of males tofemales is evident in Figure 17.2.3(b) where males typically dominate in quarters one and four but the ratio is generally more even in quarters two and three. In 2016, males dominated in all quarters, but this was within the normal range of variation which is seen for this stock over the time-series. In 2017, a more expected pattern resumed with males dominating quarters one and four. In 2018, a higher proportion of males were caught in quarters one and four but there was a more even split in quarters in two and three.

## Mean weights

The mean weights in the landings have fluctuated in this FU over the time-series. In 2018, the mean weight has declined since 2015 , and is more obvious in the trawl caught males. Mean weight for FU13 is generally lower than other areas over the time-series (Table 15.2.3). There is a trend of increasing mean weights in the samples of landings for creel catches, noticeable for both sexes, but particularly for males in the early years of the time-series. However, this has declined in recent years, although sampling levels are low, particularly in the early and most recent years of the time-series. Given the seasonal variation present in other FUs it is not possible to state with any certainty that this trend is real (Figures 15.2.4 and 15.2.5; see FU11 North Minch report, Section 12).

## Discarding

Discarding of undersized and unwanted Nephrops occurs in the Clyde fishery, and discard sampling has been conducted on the Scottish Nephrops trawler fleet since 1990. Discard rates have been high in this FU and have averaged around 27.2 \% by number in this FU since 1999. Since 2010, discard rates have been estimated to be substantially lower than the long term average and in 2017 discards decreased to a low of $9.2 \%$. In 2018, this decreased to the lowest rate in the timeseries at only $2.5 \%$ (Table 17.2.2). Studies (Charuau et al., 1982; Sangster et al., 1997; Wileman et al.,1999) suggest that some Nephropssurvive the discarding process. An estimate of $25 \%$ survival is assumed for this FU in order to calculate removals (landings + dead discards) from the population. The discard survival rate for creel caught Nephropshas been shown to behigh (ICES, 2013) and a value of $100 \%$ is used. The discard rate for use in the forecast (adjusted to account for some survival) was estimated to be $7.1 \%$ (taking a three-year average from 2016 to 2018).

## Abundance indices from UWTV surveys

Underwater TV surveys are available for both subareas since 1995 although the Sound of Jura has been surveyed more infrequently. Underwater television surveys of Nephrops burrow distributions avoid the problems associated with traditional trawl surveys that arise from variability in burrow emergence of Nephrops. TV surveys are targeted at known areas of mud, sandy mud and muddy sand in which Nephrops construct burrows. Full details of the UWTV approach can be found in the stock annex and the report of WKNEPH in 2009 (ICES, 2009). On average, 37 stations have been considered valid each year for the Firth of Clyde and 11 for the Sound of Jura. These are then raised to the estimated ground area available for Nephrops; in total $2080 \mathrm{~km}^{2}$ based on contoured superficial sediment information (British Geological Surveys). In 2019, 38 valid stations were used in the survey final analysis for the Firth of Clyde (Table17.2.3) and 12 stations for the Sound of Jura (Table17.2.4). Table17.2.5shows a detailed breakdown of information from the most recent TV surveys conducted in the Firth of Clyde. This includes estimates of abundance and variability in each of the strata adopted in the stratified random approach. Details for the Sound of Jura are shown in Table17.2.6. A CV (relative standard error) of $<20 \%$ is considered an acceptable precision level for UWTV survey estimates of abundance (SGNEPS, ICES, 2012). CVs for the three most recent TV surveys in Firth of Clyde and Sound of Jura are lower than the precision level agreed.

Figure 17.2.4 shows the distribution of stations in recent TV surveys (2014-2019) across FU13 (the two distinct subareas can be clearly seen) with the size of the symbols proportional to the Nephrops burrow density. Table 17.2.3 and Figure 17.2 .5 show the time-series estimated abundance for the TV surveys in the Firth of Clyde, with $95 \%$ confidence intervals on annual estimates. Similar information for the Sound of Jura is shown in Table 17.2.4 and Figure 17.2.6. Most surveys have shown higher density in the southern part of the Clyde. In 2018 and 2019, this appeared to still be the case.

The TV survey estimates of abundance for Nephrops in the Firth of Clyde suggest that the population increased until the mid-2000s implying a sustained period of increased recruitment. Following this, abundance has fluctuated around the values previously observed in the early 2000s. In 2019, the abundance decreased but was well within normal ranges (Figure 17.2.5).

There is not a continuous time-series of abundance in the Sound of Jura and in some years (particularly 2002 and 2006) estimates are associated with large confidence intervals. Abundance has fluctuated with no obvious trend. In 2013, the abundance was at the second lowest point in the time-series. The abundance increased in subsequent years to 2016, however decreased in 2017 and 2018 but increased in 2019 (Figure 17.2.6).

The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009; ICES, 2013). A number of potential biases were highlighted including those due to edge effects, species burrow misidentification and burrow occupancy. The cumulative relative to absolute conversion factor estimated for FU13 was 1.19 meaning that the TV survey is likely to overestimate Nephrops abundance by $19 \%$.

### 14.3 Assessment

## Comparison with previous assessments

The assessment in 2019 is based on a combination of examining trends in fishery indicators and underwater TV, using an extensive dataseries for the Firth of Clyde component of FU13 and a more limited time-series of UWTV data from the Sound of Jura subarea. The assessment in 2019 follows that of previous years (since 2015) in that the commercial data for Clyde and Sound of Jura have been combined, because of concerns regarding the accuracy of the landings data. There are also no discard samples and limited market samples available for the Sound of Jura. Therefore, the harvest rate and catches for the two areas are presented as a combined total. Nephrops abundance will continue to be monitored separately, with a TV survey in both subareas.

## State of the stock

The underwater TV surveys are presented as the best available information on the stocks of Nephrops in the two subareas of FU13. The surveys provide fishery-independent estimates of Nephrops abundance. At present, it is not possible to extract any length or age-structure information from the survey and it therefore only provides information on abundance over the area of the survey.

TV survey estimated stock abundance for the Firth of Clyde in 2019 was 2083 million individuals, a $5 \%$ decrease from the 2018 estimate but still well above the $B_{\text {trigger }}$ value of 579 million, rounded to 580 million for the provision of advice. TV survey estimated stock abundance for the Sound of Jura in 2019 was 318 million individuals, a $15.6 \%$ increase on the 2018 estimate and above the $B_{\text {trigger }}$ value of 160 million (this value does not require rounding for the provision of advice).

The calculated harvest rate for theFU13 in 2019 (dead removals for both subareas/Firth of Clyde and Sound of Jura TV abundance $=11.1 \%$ ) was below the Fmsy $^{\text {proxy }}$ value (the value associated with high long-term yield and low risk depletion) for both the Clyde (15.1\%) and the Sound of Jura ( $12.0 \%$ ). Note the FmSy proxy values for this stock was revised in October 2015 at WKMSYRef4 (ICES, 2016b).

### 14.4 Catch option table

Landings predictions and catch options at various harvest rates (based on principles established at WKNEPH (ICES, 2009)), will be made on the basis of the 2019 UWTV survey conducted in June. These will be presented in October 2019 for the provision of advice.

Catch scenario table inputs and historical estimates of mean weight in landings and harvest rates are presented in Table 17.2.2 and summarised below. The calculation of catch options for the FU13 follows the procedure outlined in the stock annex.

The table below shows the agreed inputs to the catch scenarios table.

| Input | Data | 2019 assessment |
| :--- | :--- | :---: |
| Survey abundance (millions) | UWTV 2019 | 2083 Clyde 318 SoJ |
| Mean weight in wanted <br> catch (g) | $2016-2018$ | 16.95 |
| Mean weight in unwanted <br> catch (g) | $2016-2018$ | 9.18 |
| Unwanted catch Average 2016-2018 (proportion by number; combined for Firth of <br> Clyde and Sound of Jura)  | $9.2 \%$ |  |
| Discards survival Proportion by number $2016-2018$ | $7.1 \%$ |  |

### 14.5 Reference points

Fmsy proxy for this stock was revised in October 2015 at WKMS YRef4 (ICES, 2016a; ICES, 2016b). These were updated on the basis of an average of estimated Fmsy proxy harvest rates over a period of years, which corresponds more closely to the methodology for finfish. In cases where there is a clear trend in the values a five-year average was chosen. Similarly, the five-year average of the F at $95 \%$ of the YPR obtained at the Fmsy proxy reference point was proposed as the Fmsy lower bound and the five year average of the F above FMAx that leads to YPR of $95 \%$ of the maximum as the upper bound. Using an average value also has the advantage of reducing the effect of any unusually high or low estimates of the FMSY proxy, which occasionally appear. For this functional unit the Fmsy proxy has been revised to $15.1 \%$ for the Clyde and $12.0 \%$ for the Sound of Jura respectively.
For Nephrops stocks, MSY Btrigger has been defined as the lowest stock size from which the abundancehas increased and is calculated as 579 million individuals for the Firth of Clyde. The advice from WKMS YRef4 (ICES, 2016b) rounded this value to give an MSY Btrigger of 580 million.

MSY Btrigger was not previously proposed for FU13 (SJ) as there were few points in the survey series (due to missing years). WKMS YRef4 stated that the survey series is now considered to be of sufficient length to allow the Bloss (abundance in 1995) to be proposed as the MSY Btrigger. This results in a value of 160 million (ICES, 2016b). Full details are contained in the stock annex.

These should remain under review by WGCSE and may be revised should improved data become available.

Table 17.2.2 and Figure 17.4.1 show the estimated harvest rates over this period. The harvest rate was calculated from the total dead removals for both subareas divided by the combined abundance for the Firth of Clyde TV survey and the Sound of Jura. This does result in some years were the harvest rate is not calculable as we do not have a full time-series of TV surveys for the Sound of Jura. The combined harvest rate peaked in 2007 at $43.0 \%$ before declining to around the Fmsy level for the Clyde in 2010-2011. The harvest rate has fluctuated since then and declined in 2018 to $11.1 \%$ (below $\mathrm{FMSY}^{\text {) }}$ from $17.6 \%$ in 2017 (above $\mathrm{F}_{\text {MSY }}$ ). It is unlikely that prior to 2006, the estimated harvest rates are representative of actual harvest rates due to under-reporting of landings.

### 14.6 Management strategies

Scotland has recently established a network of regional Inshore Fisheries Groups (rIFGs), nonstatutory bodies that aim to improve the management of Scotland's inshore fisheries out to six nautical miles, and to give commercial inshore fishermen a strong voice in wider marine management developments. The rIFGs will contribute to regional policies and initiatives relating to management and conservation of inshore fisheries, including impacts on the marine environment and the maintenance of sustainable fishing communities and measures designed to better conserve and sustainably exploit stocks of shellfish and sea fish (including salmon) in their local waters. Although no IFG proposals specific to the management of Nephrops fisheries have yet been adopted, some of the IFG management plans for the Scottish West Coast include spatial management of Nephrops fisheries and the introduction of creel limits.

A weekend ban on mobile gear was introduced in the Clyde in 1986 under a Scottish Statutory Instrument. Mobile gear is banned in the Inshore Clyde from Friday night to Sunday night, as are vessels greater than 21 m in length.
On the 8th of February 2016, phase 1 of the fisheries management measures for inshore MPAs in Scottish waters came into force (SG, 2016). These measures relate to both NCMPA (Marine (Scotland) Act and the UK Marine and Coastal Access Act) and Special Areas of Conservation (EC Habitats Directives - Council Directive 92/43/EEC) both of which have the aim of conserving biological diversity in Scottish waters and along with other protected sites make up Scotland's MPA network (SG,2017a). Although not specific to the management of the Nephrops fishery they will influence spatial patterns of fishing for Nephrops where controls on the two main gear types, demersal trawls and creels, are implemented on Nephrops habitat. There are three NCMPAs within the Clyde functional unit. The MPA, which extends onto the main patch of Nephrops habitat, is the South Arran NCMPA, within the Firth of Clyde subarea, where a complete ban on demersal vessels greater than 120 gross tonnage has been implemented. Partial closures (i.e. zoned management) for demersal trawlers smaller than this size and creelers are also in place. For Loch Sween, north of the main habitat area in the Sound of Jura subarea, demersal trawling by vessels is banned. However, for trawlers smaller than 75 gross tonnage, temporal closures are in place over some of the area. For the Upper Loch Fyne and Loch Goil NCMPA, just north of the main habitat area in Firth of Clyde subarea, demersal trawling by vessels greater than 75 gross tones is banned and the activity of vessels below this is zoned. Creeling activity is also zoned (SG, 2016). The areas of the NCMPAs relative to the estimated Nephrops habitat within the Clyde functional unit are presented in Figure 17.6.1.

### 14.7 Quality of assessment and forecast

There are concerns over the accuracy of historical landings and effort data and because of this the final assessment adopted is independent of official statistics. Harvest rates since 2006 are also considered more reliable due to more accurate landings data reported under new legislation.

One of the main issues for this FU is the problem of not being able to split the landings between the Sound of Jura and Firth of Clyde. This means that we are unable to provide harvest rates for the two subareas separately. What is currently provided is not actually a harvest rate for either sub area; but is likely more representative of the Firth of Clyde. This has an impact on the quality of the assessment but not on the forecast.

In recent years, the length and sex composition of the landings data is considered to be well sampled. However, in 2018 sampling levels fell below this normal standard. Discard sampling has been conducted on a quarterly basis for Scottish Nephrops trawlers in the Firth of Clyde sub-
area fishery since 1990, and is considered to represent the fishery adequately. There are few samples available from the Sound of Jura and these have been included in the FU13 raising procedure.

Underwater TV surveys have been conducted for this stock every year since 1995. The number of valid stations in the survey has remained relatively stable throughout the time period. Confidence intervals around the abundance estimates are stable throughout the series and relatively low compared with other FUs in area 6.a. In the provision of catch scenarios based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. A three-year average (2016-2018) of discard rate (adjusted to account for some survival of discarded animals) has been used in the calculation of catch advice.

The cumulative relative to absolute conversion factor estimates for FU13 component is largely based on expert opinion (see stock annex). The precision of these bias corrections cannot yet be characterised. The method to derive landings for the catch options is sensitive to the input dead discard rate and mean weight in landings and this introduces uncertainties in the catch forecasts. Precision estimates are needed for these forecast inputs.

The overall area of the ground is estimated from the available BGS contoured sediment data and at present is considered to be a minimum estimate. VMS data, recently made available and linked to landings (from queries of the Scottish FIN database) suggest no major differences between areas fished and the mud sediment maps. The inclusion of vessels smaller than 15 m would likely increase the fished area in some of the inshore locations, while in the Clyde the non-estimated sea loch areas are relatively small.

### 14.8 Recommendation for next benchmark

This stock was last benchmarked in 2009 (ICES, 2009). WGCSE recommends that the issue concerning the split of landings between Sound of Jura and the Firth of Clyde be examined when this stock is next proposed for benchmark process.

### 14.9 Management considerations

ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES division level. Management at the Functional Unit level could provide controls to ensure effort and catch were in line with resources available. In this FU, the two subareas imply that additional controls may be required to ensure that the landings taken in each subarea are in line with the landings advice.

Creel fishing takes place in part of this area although the relative scale of the fishery is smaller than in the Minches. Overall effort in terms of creel numbers is not known, and measures to control numbers are not in place. There is a need to ensure that the combined effort from all forms of fishing is taken into account when managing this stock.

There is a bycatch of other species in the area of the Firth of Clyde and estimated discards of whiting and haddock by the TR2 fleet are generally high in area 6.a. It is important that efforts continue to ensure that unwanted bycatch is kept to a minimum in this fishery. Current efforts to reduce discards and unwanted bycatches of cod include the implementation of large square meshed panels (SMPs) of 120 mm under the west coast emergency measures, and SMPs of 200 mm implemented as part of the previous Scottish Conservation Credits scheme. A seasonal closure (early spring) in the southwest part of the Firth of Clyde is in place to protect spawning cod although Nephrops vessels are derogated to fish in those parts where mud sediments are distributed.

### 14.10 References

Bailey, N. and Chapman, C. J. 1983. A comparison of density, length composition and growth of two populations off the West coast of Scotland. ICES C. M. 1983/K:42.

BGS. 2011. Marine SeaBed Sediment Map - UK Waters - 250k (DigSBS250). British Geological Survey, Nottingham. Available at: [http://www.bgs.ac.uk/discoverymetadata/13605549.html](http://www.bgs.ac.uk/discoverymetadata/13605549.html) [Accessed: unknown date].

Charuau A., Morizur Y., Rivoalen J.J. 1982. Survival of discarded Nephrops norvegicus in the Bay of Biscay and in the Celtic Sea, ICES-CM-1982/B:13.

Council Reg. (EU) 43/2009.
ESRI. 2014. ArcGIS. Version 10.2.1. Environmental Systems Research Institute, Inc.: Redlands, CA.
EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. Official Journal of the European Union, L 83: 1-17. http://data.europa.eu/eli/reg/2019/472/oj

ICES. 2010. Report of the Study Group on Nephrops Surveys (SGNEPS), 9-11 November 2010, Lisbon, Portugal. ICES CM 2010/SSGESST:22. 95 pp.

ICES. 2012. Report of the Study Group on Nephrops Surveys (SGNEPS), 6-8 March 2012, Acona, Italy. ICES CM 2012/SSGESST:19. 36 pp.
ICES. 2013. Report of the Benchmark Workshop on Nephrops assessment (WKNEPH). ICES CM 2013/ACOM:45. 230 pp.

ICES. 2016a. EU request to ICES to provide Fmsy ranges for selected stocks in ICES subareas 5 to 10. In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.2.3.1.

ICES. 2016b. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.
ICES. 2017. ICES Statistical Rectangles. International Council for the Exploration of the Sea: Copenhagen, Denmark. Available at: [http://geo.ices.dk/index.phpSource](http://geo.ices.dk/index.phpSource) [Accessed: 16/03/2017].

Sangster, G.I., Breen, M., Bova, D.J., Kynoch, R., O’Neill, F.G., Lowry. N., Moth-Poulsen, T. Hansen, U.J., Ulmestrand, M., Valentinsson, D., Hallback, H., Soldal, A.V., and Hoddevik, B. 1997. Nephrops survival after escape and discard from commercial fishing gear. Presented at ICES FTFB Working Group, Hamburg, Germany 14-17 April, 1997, ICES CM 1997 CM/B.

SG. 2016. Simple guide to fisheries management measures in Marine Protected Areas. Marine Scotland (The Scottish Government): Edinburgh. Available at: [http://www.gov.scot/Resource/0049/00498320.pdf](http://www.gov.scot/Resource/0049/00498320.pdf) [Accessed: 16/05/2017].

SG. 2017a. Marine Protected Areas in Scotland's Seas - Guidelines on the selection of MPAs and development of the MPA network. Marine Scotland (The Scottish Government): Edinburgh. Available at: [http://www.gov.scot/Topics/marine/marine-environment/mpanetwork/mpaguidelines](http://www.gov.scot/Topics/marine/marine-environment/mpanetwork/mpaguidelines) [Accessed: 16/05/2017].

SG. 2017b. Marine conservation orders (MCOs) and fisheries management measures (MPAs and SACs) with effect May 2017. Marine Scotland (The Scottish Government): Edinburgh. Available at: [https://marinescotland.atkinsgeospatial.com/nmpi/default.aspx?availablelayers=838](https://marinescotland.atkinsgeospatial.com/nmpi/default.aspx?availablelayers=838) [Accessed: 31/05/2017].

SNH. 2015. Nature Conservation Marine Protected Areas. Scottish Natural Heritage (Scottish Government): Inverness. Available at: [https://gateway.snh.gov.uk/natural-spaces/index.jsp](https://gateway.snh.gov.uk/natural-spaces/index.jsp) [Accessed: 03/04/2017].

SNH. 2016. Special Areas of Conservation. Scottish Natural Heritage (Scottish Government): Inverness. Available at: [https://gateway.snh.gov.uk/natural-spaces/index.jsp](https://gateway.snh.gov.uk/natural-spaces/index.jsp) [Accessed: 03/04/2017].
Tuck, I.D., Chapman C.J. and Atkinson, R.J.A. 1997. Population biology of the Norway lobster, Nephrops norvegicus (L.) in the Firth of Clyde, Scotland. I: Growth and density. ICES J. Mar.Sci 54, 125-135.

Tuck, I.D., Bailey, N., Atkinson, R.J.A. and Marrs, S.J. 1999. Changes in Nephrops density in the Clyde Sea area from UWTV survey data. ICES, Doc. Living Resources Comm., CM 1999/G:13 (mimeo).

Wessel, P. and Smith, W.H.F. 2016. GSHHG version 2.3.6-A Global Self-consistent, Hierarchical, Highresolution Geography Database. National Centers for Environmental Information, National Oceanic and Atmospheric Administration: Boulder, CO. [https://www.ngdc.noaa.gov/mgg/shorelines/gshhs.html](https://www.ngdc.noaa.gov/mgg/shorelines/gshhs.html) [Accessed: 31/05/2017].

Wileman, D.A., Sangster, G.I., Breen, M., Ulmestrand, M., Soldal, A.V. and Harris, R.R. 1999. Roundfish and Nephrops survival after escape from commercial fishing gear. EU Contract Final Report. EC Contract No: FAIR-CT95-0753.

Table 17.1.1. Nephrops, Cly de and Sound of Jura (FU13), ICES estimates of landings of Nephrops, 1981-2018.

| UK SCOTLAND |  |  |  |  |  | OTHER UK | IRELAND | $\begin{aligned} & \text { TO- } \\ & \text { TAL** } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | NEPHROPS TRAWL | OTHER | CREEL | BELOW MINIMUM SIZE | $\begin{aligned} & \text { SUB TO- } \\ & \text { TAL } \end{aligned}$ |  |  |  |
| 1981 | 2498 | 404 | 66 | 0 | 2968 | 0 | 0 | 2968 |
| 1982 | 2372 | 169 | 79 | 0 | 2620 | 0 | 0 | 2620 |
| 1983 | 3889 | 121 | 52 | 0 | 4062 | 14 | 0 | 4076 |
| 1984 | 3070 | 153 | 77 | 0 | 3300 | 10 | 0 | 3310 |
| 1985 | 3921 | 293 | 65 | 0 | 4279 | 7 | 0 | 4286 |
| 1986 | 4073 | 176 | 79 | 0 | 4328 | 13 | 0 | 4341 |
| 1987 | 2860 | 82 | 64 | 0 | 3006 | 3 | 0 | 3009 |
| 1988 | 3507 | 107 | 43 | 0 | 3657 | 7 | 0 | 3664 |
| 1989 | 2577 | 184 | 35 | 0 | 2796 | 16 | 0 | 2812 |
| 1990 | 2731 | 121 | 23 | 0 | 2875 | 34 | 0 | 2909 |
| 1991 | 2844 | 145 | 26 | 0 | 3015 | 23 | 0 | 3038 |
| 1992 | 2530 | 247 | 9 | 0 | 2786 | 17 | 0 | 2803 |
| 1993 | 3200 | 110 | 5 | 0 | 3315 | 28 | 0 | 3343 |
| 1994 | 2503 | 50 | 28 | 0 | 2581 | 49 | 0 | 2630 |
| 1995 | 3766 | 131 | 26 | 0 | 3923 | 64 | 0 | 3987 |
| 1996 | 3880 | 108 | 27 | 0 | 4015 | 42 | 0 | 4057 |
| 1997 | 3486 | 46 | 26 | 0 | 3558 | 63 | 0 | 3621 |
| 1998 | 4540 | 79 | 39 | 0 | 4658 | 183 | 0 | 4841 |
| 1999 | 3476 | 29 | 37 | 0 | 3542 | 210 | 0 | 3752 |
| 2000 | 3142 | 63 | 75 | 0 | 3280 | 137 | 0 | 3417 |
| 2001 | 2890 | 65 | 95 | 0 | 3050 | 132 | 0 | 3182 |
| 2002 | 3075 | 53 | 105 | 0 | 3233 | 151 | 0 | 3384 |
| 2003 | 2954 | 20 | 119 | 0 | 3093 | 80 | 0 | 3173 |
| 2004 | 2619 | 8 | 88 | 0 | 2715 | 258 | 0 | 2973 |
| 2005 | 3148 | 5 | 94 | 0 | 3247 | 148 | 0 | 3395 |
| 2006 | 4356 | 1 | 179 | 0 | 4536 | 244 | 0 | 4780 |


| UK SCOTLAND |  |  |  |  |  | OTHER UK$366$ | IRELAND$0$ | TO- <br> TAL** <br> 6660 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR <br> 2007 | NEPHROPS TRAWL$6069$ | OTHER <br> 4 | CREEL <br> 221 | BELOW MINIMUM SIZE <br> 0 | $\begin{aligned} & \text { SUB TO- } \\ & \text { TAL } \end{aligned}$ |  |  |  |
|  |  |  |  |  | 6294 |  |  |  |
| 2008 | 5320 | 3 | 184 | 0 | 5507 | 416 | 0 | 5923 |
| 2009 | 4304 | 1 | 191 | 0 | 4496 | 283 | 0 | 4779 |
| 2010 | 5162 | 5 | 211 | 0 | 5378 | 465 | 0 | 5843 |
| 2011 | 5664 | 9 | 219 | 0 | 5892 | 540 | 0 | 6432 |
| 2012 | 5617 | 4 | 203 | 0 | 5824 | 863 | 0 | 6687 |
| 2013 | 4708 | 4 | 212 | 0 | 4924 | 511 | 0 | 5435 |
| 2014 | 4770 | 1 | 258 | 0 | 5029 | 1178 | 0 | 6207 |
| 2015 | 4035 | 8 | 206 | 0 | 4249 | 898 | 0 | 5147 |
| 2016 | 4922 | 6 | 267 | 0 | 5195 | 1248 | 4 | 6447 |
| 2017 | 4021 | 3 | 256 | 0 | 4280 | 941 | 1 | 5222 |
| 2018 | 3584 | 3 | 251 | 0 | 3838 | 303 | 0 | 4141 |

Table 17.2.1. Nephrops, Cly de (FU13), ICES estimated landings of Nephrops, in each of the subareas (Firth of Clyde and Sound of Jura 1981-2018).

| YEAR | UK LANDINGS |  |  |
| :---: | :---: | :---: | :---: |
|  | FIRTH OF CLYDE | SOUND OF JURA | ALL SUBAREAS |
| 1981 | 2277 | 691 | 2968 |
| 1982 | 1983 | 637 | 2620 |
| 1983 | 3395 | 681 | 4076 |
| 1984 | 2600 | 710 | 3310 |
| 1985 | 3561 | 725 | 4286 |
| 1986 | 3228 | 1113 | 4341 |
| 1987 | 2408 | 601 | 3009 |
| 1988 | 3509 | 155 | 3664 |
| 1989 | 2595 | 217 | 2812 |
| 1990 | 2592 | 317 | 2909 |
| 1991 | 2654 | 384 | 3038 |
| 1992 | 2383 | 420 | 2803 |
| 1993 | 2766 | 577 | 3343 |
| 1994 | 2095 | 535 | 2630 |
| 1995 | 3692 | 295 | 3987 |
| 1996 | 3671 | 386 | 4057 |
| 1997 | 3135 | 486 | 3621 |
| 1998 | 4373 | 468 | 4841 |
| 1999 | 3423 | 329 | 3752 |
| 2000 | 3229 | 188 | 3417 |
| 2001 | 2979 | 203 | 3182 |
| 2002 | 3350 | 34 | 3384 |
| 2003 | 3154 | 19 | 3173 |
| 2004 | 2965 | 8 | 2973 |
| 2005 | 3388 | 7 | 3395 |
| 2006 | 4768 | 12 | 4780 |
| 2007 | 6580 | 80 | 6660 |


| YEAR | UK LANDINGS |  |  |
| :---: | :---: | :---: | :---: |
|  | FIRTH OF CLYDE | SOUND OF JURA | ALL SUBAREAS |
| 2008 | 5845 | 78 | 5923 |
| 2009 | 4688 | 91 | 4779 |
| 2010 | 5782 | 61 | 5843 |
| 2011 | 6363 | 69 | 6432 |
| 2012 | 6634 | 53 | 6687 |
| 2013 | NA | NA | 5435 |
| 2014 | NA | NA | 6207 |
| 2015 | NA | NA | 5147 |
| 2016 | NA | NA | 6443 |
| 2017 | NA | NA | 5222 |
| 2018 | NA | NA | 4141 |

Table 17.2.2. Nephrops, Clyde (FU13): Firth of Clyde and Sound of Jura combined. Adjusted TV survey abundance (Firth of Clyde subarea), landings, discard rate (proportion by number) and estimated harvest rate. The harvest rate was calculated from the total (dead) removals in number for both subareas divided by the combined abundance from both TV surveys.

| YEAR | LANDINGS IN NUMBERS (MILLIONS) | DISCARD IN NUMBERS (MILLIONS) | REMOVALS IN NUMBERS (MILLIONS)** | ADJUSTED SURVEY CLYDE (MILLIONS) | ADJUSTED SURVEY JURA (MILLIONS) | COMBINED <br> HARVEST RATE* | LANDINGS (TONNES) | DISCARDS (TONNES) | DEAD DISCARDS (TONNES) | DISCARD RATE <br> (\%) | DEAD DISCARD RATE <br> (\%) | MEAN WEIGHT IN LANDINGS (gr) | MEAN WEIGHT IN DISCARDS (gr) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 207 | 82 | 269 | 579 | 160 | 36.40 | 3987 | 619 | 464 | 28.4 | 22.90 | 19.24 | 7.54 |
| 1996 | 187 | 61 | 233 | 935 | 171 | 21.07 | 4057 | 635 | 476 | 24.7 | 19.70 | 21.68 | 10.35 |
| 1997 | 150 | 70 | 202 | 1198 | NA | NA | 3621 | 598 | 448 | 32 | 26.10 | 24.21 | 8.50 |
| 1998 | 269 | 187 | 409 | 1262 | NA | NA | 4841 | 1292 | 969 | 41 | 34.20 | 17.98 | 6.92 |
| 1999 | 216 | 93 | 286 | 930 | NA | NA | 3752 | 566 | 424 | 30.2 | 24.50 | 17.39 | 6.05 |
| 2000 | 171 | 48 | 207 | 1411 | NA | NA | 3417 | 470 | 352 | 22 | 17.40 | 19.96 | 9.75 |
| 2001 | 164 | 82 | 225 | 1486 | 272 | 12.80 | 3182 | 677 | 508 | 33.5 | 27.40 | 19.46 | 8.23 |
| 2002 | 207 | 50 | 245 | 1571 | 398 | 12.44 | 3384 | 406 | 305 | 19.5 | 15.40 | 16.35 | 8.12 |
| 2003 | 166 | 134 | 266 | 1817 | 260 | 12.81 | 3173 | 1247 | 935 | 44.7 | 37.70 | 19.13 | 9.31 |
| 2004 | 158 | 168 | 284 | 1970 | NA | NA | 2973 | 1435 | 1076 | 51.5 | 44.30 | 18.80 | 8.54 |
| 2005 | 189 | 69 | 241 | 1959 | 303 | 10.65 | 3395 | 611 | 458 | 26.8 | 21.60 | 17.96 | 8.81 |
| 2006 | 248 | 55 | 290 | 1851 | 430 | 12.71 | 4780 | 515 | 386 | 18.2 | 14.30 | 19.27 | 9.31 |
| 2007 | 350 | 387 | 640 | 1233 | 255 | 43.01 | 6660 | 2566 | 1924 | 52.5 | 45.30 | 19.05 | 6.64 |
| 2008 | 357 | 207 | 512 | 1769 | NA | NA | 5923 | 1433 | 1075 | 36.6 | 30.30 | 16.59 | 6.94 |
| 2009 | 261 | 169 | 388 | 1499 | 251 | 22.17 | 4779 | 1390 | 1043 | 39.3 | 32.70 | 18.31 | 8.23 |
| 2010 | 276 | 55 | 317 | 1750 | 376 | 14.91 | 5843 | 536 | 402 | 16.7 | 13.10 | 21.21 | 9.68 |
| 2011 | 333 | 74 | 388 | 2165 | 312 | 15.66 | 6432 | 568 | 426 | 18.2 | 14.30 | 19.34 | 7.65 |


| YEAR | LANDINGS IN NUMBERS (MILLIONS) | DISCARD IN NUMBERS (MILLIONS) | REMOVALS IN NUMBERS (MILLIONS)** | ADJUSTED SURVEY CLYDE (MILLIONS) | ADJUSTED SURVEY JURA (MILLIONS) | COMBINED HARVEST RATE* | LANDINGS (TONNES) | DISCARDS (TONNES) | DEAD DISCARDS (TONNES) | DISCARD <br> RATE <br> (\%) | DEAD DISCARD RATE <br> (\%) | MEAN WEIGHT IN LANDINGS (gr) | MEAN WEIGHT IN DISCARDS (gr) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 306 | 93 | 376 | 1421 | 371 | 20.98 | 6687 | 1066 | 800 | 23.4 | 18.60 | 21.83 | 11.42 |
| 2013 | 262 | 62 | 309 | 1990 | 198 | 14.12 | 5435 | 454 | 341 | 19 | 15.00 | 20.72 | 7.37 |
| 2014 | 295 | 78 | 353 | 1328 | 231 | 22.64 | 6207 | 696 | 522 | 20.9 | 16.60 | 20.79 | 8.92 |
| 2015 | 232 | 54 | 273 | 1820 | 376 | 12.43 | 5147 | 401 | 301 | 18.9 | 14.80 | 22.21 | 7.43 |
| 2016 | 364 | 69 | 416 | 1946 | 422 | 17.57 | 6447 | 636 | 477 | 15.9 | 12.40 | 17.70 | 9.21 |
| 2017 | 305 | 31 | 329 | 1568 | 306 | 17.56 | 5222 | 265 | 199 | 9.2 | 7.10 | 17.02 | 8.55 |
| 2018 | 268 | 7 | 273 | 2193 | 275 | 11.06 | 4141 | 68 | 51 | 2.5 | 1.90 | 16.14 | 9.79 |
| Average*** |  |  |  |  |  |  |  |  |  |  | 7.13 | 16.95 | 9.18 |

## ${ }^{*}$ Harvest rates previous to 2006 are unreliable.

** Removals numbers take the dead discard rate into account.
*** Dead discard average: 2016-2018; Mean weight in landings and discard average: 2016-2018.

Table 17.2.3. Nephrops, Clyde (FU13): Firth of Clyde subarea. Results of the 1995-2019 TV surveys (values adjusted for bias).

| YEAR | NUMBER OF VALID STATIONS | MEAN DENSITY (BURROWS / m²) | ABUNDANCE (MILLIONS) | 95\% CONFIDENCE INTERVAL (MILLIONS) |
| :---: | :---: | :---: | :---: | :---: |
| 1995 | 29 | 0.277 | 579 | 176 |
| 1996 | 38 | 0.454 | 935 | 242 |
| 1997 | 31 | 0.571 | 1198 | 262 |
| 1998 | 38 | 0.605 | 1262 | 213 |
| 1999 | 39 | 0.445 | 930 | 289 |
| 2000 | 40 | 0.681 | 1411 | 246 |
| 2001 | 39 | 0.714 | 1486 | 268 |
| 2002 | 36 | 0.756 | 1571 | 288 |
| 2003 | 37 | 0.874 | 1817 | 292 |
| 2004 | 32 | 0.95 | 1970 | 367 |
| 2005 | 44 | 0.941 | 1959 | 287 |
| 2006 | 43 | 0.882 | 1851 | 257 |
| 2007 | 40 | 0.597 | 1233 | 218 |
| 2008 | 38 | 0.849 | 1769 | 291 |
| 2009 | 39 | 0.723 | 1499 | 210 |
| 2010 | 37 | 0.84 | 1750 | 327 |
| 2011 | 40 | 1.041 | 2165 | 305 |
| 2012 | 37 | 0.681 | 1421 | 227 |
| 2013 | 34 | 0.956 | 1990 | 246 |
| 2014 | 35 | 0.639 | 1328 | 237 |
| 2015 | 37 | 0.875 | 1820 | 351 |
| 2016 | 37 | 0.935 | 1946 | 249 |
| 2017 | 38 | 0.754 | 1568 | 239 |
| 2018 | 40 | 1.055 | 2193 | 297 |
| 2019 | 38 | 1.002 | 2083 | 381 |

Table 17.2.4. Nephrops, Clyde (FU13): Sound of Jura subarea. Results of the 1995-2019 TV surveys (values adjusted for bias).

| YEAR | NUMBER OF VALID STATIONS | MEAN DENSITY (BURROWS / m²) | ABUNDANCE (millions) | 95\% CONFIDENCE INTERVAL (millions) |
| :---: | :---: | :---: | :---: | :---: |
| 1995 | 7 | 0.42 | 160 | 58 |
| 1996 | 10 | 0.45 | 171 | 26 |
| 1997 | no surveys |  |  |  |
| 1998 |  |  |  |  |
| 1999 |  |  |  |  |
| 2000 |  |  |  |  |
| 2001 | 13 | 0.71 | 272 | 76 |
| 2002 | 9 | 1.04 | 398 | 167 |
| 2003 | 12 | 0.68 | 260 | 68 |
| 2004 | no survey |  |  |  |
| 2005 | 11 | 0.79 | 303 | 84 |
| 2006 | 10 | 1.13 | 430 | 134 |
| 2007 | 10 | 0.67 | 255 | 58 |
| 2008 | no survey |  |  |  |
| 2009 | 12 | 0.66 | 251 | 68 |
| 2010 | 12 | 0.98 | 376 | 39 |
| 2011 | 12 | 0.82 | 312 | 73 |
| 2012 | 12 | 0.98 | 371 | 61 |
| 2013 | 9 | 0.52 | 198 | 35 |
| 2014 | 9 | 0.61 | 231 | 90 |
| 2015 | 12 | 0.98 | 376 | 127 |
| 2016 | 12 | 1.11 | 422 | 42 |
| 2017 | 12 | 0.80 | 306 | 71 |
| 2018 | 12 | 0.72 | 275 | 53 |
| 2019 | 12 | 0.832 | 318 | 61 |

Table 17.2.5. Nephrops, Clyde (FU13): Firth of Cly de subarea. Results by stratum of the 2017-2019 TV surveys. Note that stratification was based on a series of sediment strata (M - Mud, SM - Sandy mud, MS - Muddy sand).

| STRATUM | AREA <br> (km ${ }^{2}$ ) | NUMBER OF STATIONS | MEAN BURROW DENSITY (no./m) | OB- <br> SERVED <br> VARI- <br> ANCE | ABUN- <br> DANCE <br> (MILLIONS) | STRATUM <br> VARI- <br> ANCE | PROPORTION OF TOTAL VARIANCE | SURVEY <br> PRECISION <br> LEVEL <br> (RSE) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 TV survey |  |  |  |  |  |  |  |  |
| M | 716.8 | 14 | 0.634 | 0.048 | 454.8 | 1761 | 0.123 |  |
| SM | 698.6 | 12 | 0.938 | 0.201 | 655 | 8163 | 0.57 |  |
| MS | 664.6 | 12 | 0.69 | 0.119 | 458.4 | 4387 | 0.307 |  |
| Total | 2080 | 38 |  |  | 1568.2 | 14312 | 1 | 0.07 |
| 2018 TV survey |  |  |  |  |  |  |  |  |
| M | 716.8 | 13 | 1.015 | 0.122 | 727.5 | 4821 | 0.219 |  |
| SM | 698.6 | 14 | 1.284 | 0.285 | 897 | 9909 | 0.45 |  |
| MS | 664.6 | 13 | 0.855 | 0.215 | 568.8 | 7312 | 0.332 |  |
| Total | 2080 | 40 |  |  | 2193.3 | 22042 | 1.001 | 0.07 |
| 2019 TV survey |  |  |  |  |  |  |  |  |
| M | 716.8 | 14 | 0.841 | 0.096 | 602.8 | 3517 | 0.097 |  |
| SM | 698.6 | 11 | 1.329 | 0.458 | 928.1 | 20296 | 0.559 |  |
| MS | 664.6 | 13 | 0.831 | 0.367 | 552.4 | 12467 | 0.344 |  |
| Total | 2080 | 38 |  |  | 2083.3 | 36279 | 1 | 0.09 |

Table 17.2.6. Nephrops, Clyde (FU13): Sound of Jura subarea. Results by stratum of the 2017-2019 TV surveys. Note that stratification was based on a series of sediment strata.

| STRA- <br> TUM | AREA (km2) | NUMBER OF STATIONS | MEAN BURROW DENSITY (no./m2) | OBSERVED VARIANCE | ABUN- <br> DANCE (MIL- <br> LIONS) | STRATUM VARIANCE | PROPORTION OF TOTAL VARIANCE | SURVEY <br> PRECISION <br> LEVEL SUR- <br> VEY (RSE) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2017 \text { TV }$ <br> survey |  |  |  |  |  |  |  |  |
| M | 90 | 2 | 0.454 | 0.014 | 40.8 | 57 | 0.045 |  |
| SM | 150 | 5 | 0.879 | 0.124 | 131.8 | 560 | 0.44 |  |
| MS | 142 | 5 | 0.939 | 0.163 | 133.4 | 657 | 0.515 |  |
| Total | 382 | 12 |  |  | 306.1 | 1274 | 1 | 0.073 |
| $2018 \text { TV }$ <br> survey |  |  |  |  |  |  |  |  |
| M | 90 | 3 | 0.739 | 0.019 | 66.6 | 52 | 0.075 |  |
| SM | 150 | 4 | 0.691 | 0.008 | 103.7 | 43 | 0.062 |  |
| MS | 142 | 5 | 0.734 | 0.148 | 104.3 | 598 | 0.863 |  |
| Total | 382 | 12 |  |  | 274.5 | 693 | 1 | 0.10 |
| $2019 \text { TV }$ <br> survey |  |  |  |  |  |  |  |  |
| M | 90 | 2 | 0.689 | 0.088 | 62 | 357 | 0.389 |  |
| SM | 150 | 4 | 0.878 | 0.023 | 131.8 | 128 | 0.139 |  |
| MS | 142 | 6 | 0.874 | 0.129 | 124.1 | 434 | 0.472 |  |
| Total | 382 | 12 |  |  | 317.9 | 919 | 1 | 0.101 |

$\qquad$

## Landings - International



Effort - Scottish Nephrops trawlers


Figure 17.2.1. Nephrops, Cly de (FU13). Long-term landings and effort.

## Length frequencies for catch (dotted) and landed(solid): Nephrops in FU13



Figure 17.2.2. Nephrops, Clyde (FU13). Catch length-frequency distribution and mean size in catches (dotted) and landings (solid) for Nephrops, 2000-2018.. Vertical lines are minimum conservation reference size ( 25 mm ) and 35 mm .

## Landings




Figure 17.2.3.(a) Nephrops, Cly de (FU13). Landings by quarter and sex from Scottish trawlers.


Figure 17.2.3. (b) Nephrops, Cly de (FU13), Proportion of males by quarter (1980-2018).


Figure 17.2.4. Nephrops, Clyde (FU13), TV survey station distribution and relative density (burrows/m²) for Firth of Clyde and Sound of Jura subareas, 2014-2019. Sound of Jura located to the east. Shaded green and brown areas represent areas of suitable sediment for Nephrops. Bubbles scaled the same. Red crosses represent zero observations.

## Firth of Clyde



Figure 17.2.5. Nephrops, Clyde (FU13): Firth of Clyde subarea. Time-series of revised TV survey abundance estimates (adjusted for bias), with $95 \%$ confidence intervals, 1995-2019. The dashed blue line is the rounded $B_{\text {triger }}$ value of 580 million individuals.

## Sound of Jura



Figure 17.2.6. Nephrops, Cly de (FU13): Sound of Jura subarea. Time-series of TV survey abundance estimates (adjusted for bias) with $95 \%$ confidence intervals, 1995-2019. The dashed blue line is the rounded $B_{\text {triger }}$ value of $\mathbf{1 6 0}$ million individuals.

Harvest Rate


Figure 17.4.1. Cly de (FU13) Nephrops harvest rate, 1995-2019. The harvest rate is calculated by dead removals (both subareas combined)/TV abundances (both sub-areas combined). The dashed and solid lines are the FMSY proxy harvest rate (for the Firth of Clyde $\mathbf{1 5 . 1 \%}$ ) and the harvest rate respectively. Harvest rates prior to 2006 are unreliable.


Figure 17.6.1. The area of Nephrops habitat (Mud, Muddy Sand and Sandy Mud) within the Clyde functional unit (FU13) relative to the areas of the Nature Conservation MPAs (NCMPAs) which fisheries management measures. Areas where demersal trawling is prohibited, restricted (i.e. vessel size restrictions or seasonal closures) and where creeling is prohibited are displayed. For more detailed information see SG (2016). Geographic Coordinate System: OSGB 1936, Datum: OSGB 1936, Projected Coordinate System: British National Grid. Coastline by Wessel and Smith (2016), MPA sites subsetted from NCMPA (SNH, 2015) and SAC (SNH, 2016) layers, management areas by SG (2017b) and functional units generated from merged ICES rectangles (ICES, 2017). Map and modified layers created using ArcGIS (ESRI, 2014).

## 15 Norway lobster (Nephrops norvegicus) in Division 7.a, Functional Unit 14 (Irish Sea East)

### 15.1 Nephrops Subarea 7 general section

## Stock description and management units

A TAC is in place for ICES Area 7, which does not correspond to the assessment units. As Nephrops are limited to muddy habitats, the distribution of suitable sediment defines the species distribution and the stocks are therefore assessed as eight separate Functional Units. There are also some smaller catches from areas outside these Functional Units. The ICES statistical rectangles covered by the Functional Units in ICES Area 7 are listed in the table below.

| FU no. | Name | ICES <br> Divisions | ICES Statistical rectangles |
| :---: | :---: | :---: | :---: |
| 14 | Irish Sea East | 7 a | 35-38E6; 38E5 |
| 15 | Irish Sea West | 7 a | 35E3, 36E3; 35-37 E4-E5; 38E4 |
| 16 | Porcupine Bank | 7b, c,j, ${ }^{\text {c }}$ | 31-35 D5-D6; 32-35 D7-D8 |
| 17 | Aran Grounds | 7b | 34-35 D9-E0 |
| 18* | Northwest Irish Coast | 7b | 36-37 D9; 37E0-E1 |
| 19 | Southeast and southwest Irish Coast | 7a,g,j | 31-33 D9-E0; 31E1; 32E1-E2; 33E2-E3 |
| 20-21 | Labadie, Jones and Cockburn Bank | 7g,h | 28 EO-E2; 29 E0-E3; 30E1-E3; 31E2 |
| 22 | Smalls Ground | 7g | 31-32 E3-E4 |

* Landings from FU18 are reported to other statistical rectangles outside FUs as these are minimal. WGCSE will monitor FU18 landings in case of any fishery developments.

Nephrops Functional Units in Subarea 7 (FU 14-22). The TAC covers all of Subarea 7. (Note: Functional Units in Subarea 6 (FU 11-13) also shown):


## Landings Obligation

On the West Coast and around Ireland (FU 11-22), in 2017, vessels where 30\% or more of their landings in 2014 and 2015 were Nephrops had to land all Nephrops. In 2018, vessels where 20\% or more of their landings in 2015 and 2016 were Nephrops had to land all Nephrops. High survival exemptions existed for creel caught Nephrops. De minimis exemptions apply to Nephrops vessels, for Subarea 7 allowing them to discard Nephrops, as long as they made up no more than $7 \%$ of the catch in 2016 and 2017; this decreased to $6 \%$ in 2018 and will do again to $5 \%$ for 2019.

## Minimum Conservation Reference Size (Minimum landing size)

Under the Landing Obligation, minimum landings sizes have been abolished. Instead a Minimum Conservation Reference Size(MCRS) for each specieshas been introduced. Unless exempt, Nephrops below the MCRS must be landed and may be sold but cannot go for human consumption. In most cases, the MCRS is the same as old MLS, being 25 mm carapace length (or over 85 mm total length) around Ireland (FUs 16-22); the MCRS is $20 \mathrm{~mm} \mathrm{CL}(>70 \mathrm{~mm} \mathrm{TL}$ ) on the West coast (6.a, FUs 11-13) and the Irish Sea (7a, FUs 14-15).

The MCRS implemented for the Irish Sea at 20 mm CL is less than the rest of the ICES Area 7 (set at $25 \mathrm{~mm} C L$ ) and applies to the Irish and UK fleets. A more restrictive regulation is adopted by the French Producers' Organisations ( 35 mm CL or 115 mm TL ) to all French trawlers.

## Management applicable in 2017 and 2018

The TAC is currently set for the whole Area 7. The TAC for 2018 was 29091 t , this represented an increase of $15 \%$ in relation to 2017 with 25356 t . The TAC area includes a number of Nephrops stocks showing different levels of exploitation. A single TAC covering a number of distinct stocks allows the possibility of unrestricted catches being taken from a heavily exploited stock when advice suggests they should be limited.

Details of all regulations including effort controls in place are provided in the stock annex for all functional units under this subarea.

Council Regulation (EU) 2018/120 of 23 January 2018 fixing for 2018 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters, and amending Regulation (EL) 2017/127.

TAC in 2018

| Species: | Norway lobster <br> Nephrops norvegicus | Zone: | 7 <br> (NEP/07.) |
| :--- | :--- | :--- | :--- |
| Spain |  | 1745 |  |
| France | 7074 |  |  |
| Ireland | 10729 |  |  |
| United Kingdom | 9543 |  |  |
| Union | 29091 | Analytical TAC |  |
| TAC | 29091 | Arrticle 12(1) of this Regulation applies |  |

Special condition:
Within the limits of the abovementioned quotas, no more than the quantities given below may be taken in the following zone:

|  | Functional Unit $\mathbf{1 6}$ of ICES Subarea <br> 7 <br> $7(N E P / * 07 U 16):$ |
| :--- | ---: |
| Spain | 825 |
| France | 516 |
| Ireland | 992 |
| United Kingdom | 401 |
| Union | 2734 |

COUNCIL REGULATION (EU) 2017/127 of 20 January 2017 fixing for 2017 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters.

TAC in 2017

| Species: | Norway lobster <br> Nephrops norvegicus | Zone: | VII <br> (NEP/07.) |
| :--- | :--- | :--- | :--- |
| Spain |  | 1521 |  |
| France | 6166 |  |  |
| Ireland | 9352 |  |  |
| United Kingdom | 8317 |  |  |
| Union | 25356 | Analytical TAC |  |
| TAC | 25356 | Article 11(1) of this Regulation applies |  |

Special condition:
within the limits of the abovementioned quotas, no more than the quantities given below may be taken in the following zone:

|  | Functional Unit 16 of ICES Subarea <br> VII (NEP/*07U16): |
| :--- | ---: | ---: |
| Spain | 935 |
| France | 586 |
| Ireland | 1124 |
| United Kingdom | 455 |
| Union | 3100 |

## Landings area 7

Table below gives the summary of reported landings by Functional Unit for ICES Area 7.

| Year | FU 14 - <br> Irish <br> Sea <br> East | FU 15 - <br> Irish Sea <br> West | FU 16 - Porcupine Bank | FU 17 - <br> Aran <br> Grounds | *FU 18 - Ireland Northwest Coast | FU 19 - Ireland Southwest and Southeast coast | FU 20-21 - <br> Labadie, Jones, Cockburn | FU 22 - <br> Smalls <br> Grounds | FUs 20+21+22All Celtic Sea FUs combined | Other statistical rectangles Outside FUs | Total Landings ICES Subarea 7 | TAC for 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 961 | 7,296 | 1,744 | 481 |  |  |  |  |  | 249 | 10731 |  |
| 1979 | 900 | 8,948 | 2,269 | 452 |  |  |  |  |  | 237 | 12806 |  |
| 1980 | 730 | 4,578 | 2,925 | 442 |  |  |  |  |  | 205 | 8880 |  |
| 1981 | 829 | 7,249 | 3,381 | 414 |  |  |  |  |  | 382 | 12255 |  |
| 1982 | 869 | 9,315 | 4,289 | 210 |  |  |  |  |  | 234 | 14917 |  |
| 1983 | 763 | 9,448 | 3,426 | 131 |  |  |  |  | 3,667 | 174 | 17609 |  |
| 1984 | 602 | 7,760 | 3,571 | 324 |  |  |  |  | 3,653 | 187 | 16097 |  |
| 1985 | 498 | 6,901 | 3,919 | 207 |  |  |  |  | 3,599 | 194 | 15318 |  |
| 1986 | 671 | 9,978 | 2,591 | 147 |  |  |  |  | 2,638 | 113 | 16138 |  |
| 1987 | 449 | 9,753 | 2,499 | 62 |  |  |  |  | 3,409 | 107 | 16279 | 24,700 |
| 1988 | 462 | 8,586 | 2,375 | 828 |  |  |  |  | 3,165 | 140 | 15556 | 24,700 |


| Year | FU 14 - <br> Irish <br> Sea <br> East | FU 15 - <br> Irish Sea <br> West | FU 16 - Porcupine Bank | FU 17 - <br> Aran <br> Grounds | *FU 18 - Ireland Northwest Coast | FU 19 - Ireland Southwest and Southeast coast | FU 20-21 - <br> Labadie, Jones, Cockburn | FU 22 - <br> Smalls <br> Grounds | FUs 20+21+22- <br> All Celtic Sea FUs combined | Other statistical rectangles Outside FUs | Total Landings ICES Subarea 7 | TAC for 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 401 | 8,128 | 2,115 | 344 |  | 899 |  |  | 4,005 | 134 | 16026 | 26,000 |
| 1990 | 563 | 8,300 | 1,895 | 519 |  | 754 |  |  | 4,290 | 102 | 16423 | 26,000 |
| 1991 | 747 | 9,554 | 1,640 | 410 |  | 1077 |  |  | 3,295 | 169 | 16892 | 26,000 |
| 1992 | 427 | 7,541 | 2,015 | 372 |  | 888 |  |  | 4,165 | 409 | 15817 | 20,000 |
| 1993 | 515 | 8,102 | 1,857 | 372 | 10 | 905 | 3,466 | 1182 |  | 455 | 16864 | 20,000 |
| 1994 | 447 | 7,606 | 2,512 | 729 | 126 | 390 | 4202 | 941 |  | 570 | 17523 | 20,000 |
| 1995 | 584 | 7,796 | 2,936 | 866 | 26 | 695 | 3536 | 1081 |  | 397 | 17917 | 23,000 |
| 1996 | 475 | 7,247 | 2,230 | 525 | 46 | 888 | 2822 | 937 |  | 623 | 15793 | 23,000 |
| 1997 | 566 | 9,971 | 2,409 | 841 | 15 | 756 | 2038 | 944 |  | 340 | 17880 | 23,000 |
| 1998 | 388 | 9,128 | 2,155 | 1,410 | 78 | 827 | 1713 | 835 |  | 514 | 17048 | 23,000 |
| 1999 | 624 | 10,786 | 2,289 | 1,140 | 16 | 579 | 1,152 | 1775 |  | 322 | 18683 | 23,000 |
| 2000 | 567 | 8,370 | 910 | 880 | 9 | 696 | 1,778 | 2890 |  | 243 | 16344 | 21,000 |
| 2001 | 532 | 7,441 | 1,222 | 913 | 2 | 815 | 1,833 | 2938 |  | 368 | 16064 | 18,900 |


| Year | FU 14 - <br> Irish <br> Sea <br> East | FU 15 - <br> Irish Sea West | FU 16 - Porcupine Bank | FU 17 - <br> Aran <br> Grounds | *FU 18 - Ireland Northwest Coast | FU 19 - Ireland Southwest and Southeast coast | FU 20-21 - <br> Labadie, Jones, Cockburn | FU 22 - <br> Smalls <br> Grounds | FUs 20+21+22All Celtic Sea FUs combined | Other statistical rectangles Outside FUs | Total Landings ICES Subarea 7 | TAC for 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 577 | 6,793 | 1,327 | 1,154 | 14 | 1318 | 2,674 | 1993 |  | 243 | 16093 | 17,790 |
| 2003 | 376 | 7,052 | 1,064 | 933 | 16 | 1239 | 2,953 | 2065 |  | 186 | 15884 | 17,790 |
| 2004 | 472 | 7,266 | 1,406 | 525 | 22 | 1074 | 2,443 | 1828 |  | 161 | 15197 | 17,450 |
| 2005 | 570 | 6,529 | 2,197 | 778 | 15 | 711 | 2,469 | 2533 |  | 180 | 15982 | 19,544 |
| 2006 | 628 | 7,535 | 2,185 | 637 | 14 | 741 | 2,523 | 1761 |  | 270 | 16294 | 21,498 |
| 2007 | 959 | 8,424 | 2,074 | 913 | 3 | 957 | 2,419 | 2950 |  | 206 | 18905 | 25,153 |
| 2008 | 726 | 10,482 | 1,000 | 1,057 | 1 | 841 | 2,980 | 3090 |  | 111 | 20288 | 25,153 |
| 2009 | 693 | 9,166 | 879 | 626 | 10 | 833 | 3,145 | 2185 |  | 322 | 17860 | 24,650 |
| 2010 | 583 | 8,929 | 922 | 939 | 7 | 722 | 1,793 | 2714 |  | 316 | 16925 | 22,432 |
| 2011 | 561 | 10,159 | 1,278 | 659 | 13 | 608 | 1,237 | 1636 |  | 359 | 16510 | 21,759 |
| 2012 | 531 | 10,527 | 1,258 | 1,246 | 28 | 770 | 1,189 | 2618 |  | 110 | 18276 | 21,759 |
| 2013 | 495 | 8,672 | 1,141 | 1,295 | 0 | 781 | 1,387 | 2257 |  | 325 | 16354 | 23,605 |
| 2014 | 679 | 8,613 | 1,189 | 766 | 0 | 468 | 1,836 | 2526 |  | 194 | 16271 | 20,989 |


| Year | FU 14 - <br> Irish <br> Sea <br> East | FU 15 - <br> Irish Sea <br> West | FU 16 - Porcupine Bank | FU 17 - <br> Aran <br> Grounds | *FU 18 - Ireland Northwest Coast | FU 19 - Ireland Southwest and Southeast coast | FU 20-21 - <br> Labadie, Jones, Cockburn | FU 22 - <br> Smalls <br> Grounds | FUs 20+21+22- <br> All Celtic Sea FUs combined | Other statistical rectangles Outside FUs | Total Landings ICES Subarea 7 | TAC for 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 378 | 8,632 | 1,394 | 370 | 0 | 507 | 2,116 | 2350 |  | 174 | 15921 | 21,619 |
| 2016 | 237 | 7327 | 2154 | 641 | 0 | 591 | 2,453 | 3329 |  | 80 | 16812 | 23348 |
| 2017 | 265 | 6149 | 2632 | 295 | 0 | 420 | 1,849 | 3560 |  | 137 | 15307 | 25356 |
| 2018 | 268 |  |  |  |  |  |  |  |  |  |  | 29091 |
| Average | 530 | 8301 | 2081 | 648 | 19 | 784 | 2,504 | 2,117 | 4,621 | 231 | 16210 | 22,540 |

${ }^{*}$ Landings from FU18 are reported to other statistical rectangles outside FUs as these are minimal since 2013. WGCSE will monitor FU18 landings in case of any fishery developments.

## Nephrops FU14 section

## Type of assessment in 2019

This stock was inter-benchmarked in September 2015 (ICES, 2015) and the assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the inter-benchmark process and described in the stock annex (updated at WGCSE 2018). The UWTV survey undertaken in the summer 2019 forms the basis of advice for this stock.

## ICES advice applicable to 2018

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2014-2016, catches in 2018 should be no more than 1281 tonnes.

To ensure that the stock in Functional Unit 14 is exploited sustainably, management should be implemented at the functional unit level."

## ICES advice applicable to 2019

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2016-2017, catches in 2019 should be no more than 922 tonnes.

To ensure that the stock in Functional Unit 14 is exploited sustainably, management should be implemented at the functional unit level."

### 15.2 General

## Stock description and management units

The Irish Sea East Nephrops stock (FU14) is in ICES Subarea 7and comprises ICES rectangles: 38E5,38E6,37E6,36E6,35E6.

In FU14 Nephrops are caught on two spatially discrete grounds. Most of the fishery takes place on the main ground located between the West coast of England and Isle of Man, additionally there is also fishing activity in a small inshore ground known as Wigtown Bay.


East Irish Sea fishing grounds: A= Main fishing ground; B= Wigtown bay area. Windfarms represented by red polygons. (Source: ICES, 2015).

Main landing ports: Whitehaven, Fleetwood, Mary port and Kilkeel.

## Fishery in 2018

The Eastern Irish Sea Nephropsfishery is dominated by UK activity, representing on average 93\% of the reported annual international landings (2008-2018). This is a relative small fishery compared to other FUs in the TAC area. Landings have been generally declining over the past eleven years (Table3.8.2), from a high of 959 tonnes in 2007 to a low of 237 tonnes in 2016, with landings of 268 tonnes in 2018. The main fleets targeting Nephrops include directed single-rig and twin-rig otter trawlers operating out of ports in UK (E\&W), UK (NI), Republic of Ireland and UK (S).

As in previous years, in 2018, the UK fleet accounted for the highest proportion of landings in tonnes.

A more detailed historical fishery description is provided in the stock annex.

## Information from stakeholders

No additional information was provided.

### 15.3 Data

## InterCatch

Data for 2018 were successfully uploaded into InterCatch prior the 2019 WG meeting. Uploaded data were worked up in InterCatch to generate 2018 raised international length-frequency distributions, and to derive catch and discard length frequencies for 2018. The same allocation procedure was followed as in 2017: English landings were raised with English samples, Northern Irish landings raised with Northern Irish samples and all other remaining landings with pooled English and Northern Irish samples.

## Landings

Official landings as reported to ICES from FU14 are presented in Table 3.8.1.
There are reported landings for this functional unit since 1973 with a minimum and maximum of 178.7 t (in 1974) and 960.5 t (in 1978) respectively. Between 1987 and 2006, landings from FU14 appeared relatively stable fluctuating around a long-term average of about 550 t . The introduction of the Buyers and Sellers legislation in 2006 by the UK precludes direct comparison with previous years, as reported levels are considered to have significantly improved. Over the period 2007-2018, landings have declined considerably from the peak year of 2007; landings in 2018 (268 t) were very similar to 2017 (265t).

## Effort

Following discussions at WGCSE, it was concluded that effort should be reported in the WGCSE report in kWdays , and LPUE should be reported in KG/kWdays in the knowledge that the trend is likely to be a biased underestimate because it is not adjusted for efficiency or behavioural changes. Thetime-series of effort and LPUE is updated in Table3.8.3 and Figure 3.8.2. There has been a general decline in targeted effort across the available time series and is now around the lowest recorded levels (although 2018 did see a small rise in effort compared to 2017).

## Sampling levels

Sampling levels, data aggregating and raising procedures were reviewed by IBPNeph 2015, documented in the stock annex, and examined further at WGCSE 2018. Recent sampling levels have fluctuated; prior to 2016 sample data have only been available from landings into England; however, since 2016 samples have also been available from landings into Northern Ireland. In 2018, nine English samples (four from at-sea observations and five from port-side catch sampling) and two Northern Irish samples were used to raise the data.

## Commercial length-frequency distributions

The raised catch length distributions are shown in Figure 3.8.3. The mean sizes for both sexes from 2008 fluctuate considerably. For 2018, the mean size of the landings was higher than the record low of 2016.

## Length composition

Between 2010 and 2012, sampling levels are considered insufficient to reliably characterise the length composition of extractions. Increased sampling levels from 2013 onwards have allowed for length compositions to be constructed. For 2013 and 2014, a full revision was done through an inter-benchmark process (ICES, 2015; described in the stock annex). Data aggregating and raising procedures from 2015 to 2017 were conducted according tobenchmark procedures (ICES, 2005) and referred in the stock annex. These were revised during WGCSE 2018 to account for Northern Irish sampling data since 2016 and are described further in the stock annex.

Updated historical trends in length distributions and proportion discarded are shown in Figure 3.8.3 and Table 3.8.4. Discard selection curve estimates for the East Irish Sea shows a L50=23.54 and a L25=24.77 mm CL (Figure 3.8.4), which shows a selectivity at higher sizes compared with FU15.

Mature females are mainly caught in the non-berried state between the moulting, which reaches its peak in May. Females mature at about 23 mm carapace length. (Thomas and José Figueiredo, 1965).

## Sex ratio

The catch sex ratio by year is shown in Figure 3.8.5. The ratio is quite variable but average sex ratio is $54 \%$ male (1999-2018), the sex ratio for 2018 being just above this ( $58 \%$ ).

## Mean weight explorations

The annual mean weight estimate for landings and discards is provided in Table 3.8.4 and in Figure 3.8.6. There is a substantial difference between the mean weights prior to 2011 and after 2013 (the gap being where sampling was too low to be reliable). Since 2016, NI sampling has been included, and the mean weight of NI samples is considerably lower than for English sampling (e.g. in 2018, mean weight of landings from English sampling was 30.7 g compared to 21.5 in Northern Irish sampling). As a result, comparison with years prior to 2016 is not practical. Mean weights over the last three years (2016-2018) are variable without trend.

## Discarding

Discard selection was revised at the IBP process in 2015 (ICES, 2015) and described in the stock annex. Figure 3.8 .4 shows a single discard ogive fitted by pooling all years (2003-2014) and mesh sizes. Final discard selection for the East Irish Sea shows a L50 $=23.54$ and a L25 $=24.77 \mathrm{~mm}$ CL (Figure 4.3.4), which shows a selectivity at higher sizes compared with FU15. Due to high interannual variation in mean sizes of both landings and discards, the discard ogive was not updated using 2015 to 2018 data.

Table 3.8.5 gives raised international landings and discard weight and numbers by year.
At IBPNeph (ICES, 2015), it was agreed that the discard survival rate should be up-dated from $0 \%$ to $10 \%$. Although there are no direct survivability studies available for this area, it is expected that the survivability of discarded animals should be similar to the fishery in FU15 where fishing practices are similar, and both are largely spring/summer fisheries and animals discarded are exposed to warmer temperatures before being returned to the sea.

## Abundance indices from UWTV surveys

In August of 2007-2019 the UKand the Republic of Ireland carried out an underwater TV survey of the Nephrops grounds in theeastern IrishSea. The survey is of a fixed grid design and is carried out using the same protocols used in UWTV surveys in the western Irish Sea (ICES, 2007; ICES, 2014). The survey stations used in 2019 are presented in Figure 3.8.7.

Due to the construction of the windfarm in the southern part of the ground, the survey area was reviewed at IBP 2015 but the protocols and standardised process to run the survey were not modified (see stock annex and IBP 2015 report ICES, 2015). The new survey area (based on a cokriging model) is shown in Figure 3.8.8. The boundary used to define the ground limits for absolute abundance runs close to the outer survey stations.

| Ground | Area Km $^{\mathbf{2}}$ | Source |
| :--- | :---: | :--- |
| Main ground 2008-2010 | 1032.75 | WGCSE 2008 |
| Main ground 2011-2019 | 1019.79 | IBP 2015 - ICES, 2015 |
| Wigtown Bay | 67.21 | IBP 2015 - ICES, 2015 |

Wigtown Bay in relation to Main ground $=6.6 \%$ * (increase from $1.9 \%$ prior to the windfarm construction).

Abundance indexes were revised back to 2011, the year where the effect of effort displacement is clearly visible due to the windfarm construction. Final updated abundance burrow density estimates are presented in Table 3.8.6 and visualised in Figure 3.8.9 where the geo-spatial model was updated using the new area based on the co-kriging approach ( $1019.79 \mathrm{Km}^{2}$ ) and theextrapolation to Wigtown Bay using 6.6\%.

The abundance estimate for 2019 ( 399 million) is a decrease of $22 \%$ compared to the 2018 figure of 514 million (Figure 3.8.10). This is the second lowest value in the series, but is only $18 \%$ below the 2008-2018 average. The surveys show a clear spatial distribution pattern, with highest densities in the central north of the patch and more variable in the area further south. The grounds are fairly well delineated by consistently low-density ground to the northeast and west (Figure 3.8.9). CV s over the entire time-series (Table 3.8.6) are within the accepted precision level of 20\% (ICES, 2012).

The use of the UWTV surveys for the provision of Nephrops management advice was extensively reviewed by WKNEPH (2009). A number of potential factors were highlighted, including those due to edge effects; species burrow misidentification and burrow occupancy. Using the same process adopted at WKNEPH, a cumulative absolute conversion factor for this FU was predicted to be 1.2 for FU14 (see stock annex) which means the TV survey is likely to overestimate Nephrops abundance by $20 \%$. The burrow abundances shown in Table 3.8.5 and Figure 3.8.9 have been adjusted using this conversion factor since 2008.

### 15.4 Assessment

## Comparison with previous assessments

The WGCSE 2019 carried out an UWTV based assessment for this stock. The methods used were very much in line with WKNEPH (ICES, 2009) and the approach taken for other Nephrops stocks in 6 and 7 by WGCSE. This approach was inter-benchmarked at IBPNeph (ICES,2015). The 2019 assessment uses a three-year average for weights and discarding rates (2016-2018). This is in contrast to the 2018 assessment, which only had reliable weights for both fleet components (England and Northern Ireland) for 2016-2017.

## State of the stock

UWTV abundance estimates suggest that the stock size has fluctuated between abundance values of 350 and 694 million Nephrops. The 2019 estimate ( 399 million) decreased by $22 \%$ in relation to 2018 and although below average is above the MSY $\mathrm{B}_{\text {trigger }}(350$ million).

Table 3.8 .5 and Figure 3.8 .11 summarise the abundance estimated including the confidence intervals and the harvest ratios (\% dead removed / UWTV abundance) which have been above the


### 15.5 Catch scenarios table

Catch scenarios table inputs and historical estimates of mean weight in landings and harvest ratios are presented in Table 3.8.5 and summarised below. The calculation of catch options for the FU14 follows the procedure outlined in the stock annex. The basis for the catch options:

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Stock abundance | 399 million | UWTV 2019 |
| Mean weight in wanted catch | 18.72 g | Average 2016-2018 |
| Mean weight in unwanted <br> catch | 9.22 g | Average 2016-2018 |
| Unwanted catch proportion | $12.71 \%$ | Average 2016-2018 (proportion by number) |
| Unwanted catch survival rate | $10 \%$ | Only applies inscenarios where discarding is allowed. |
| Dead unwanted catch rate | $11.59 \%$ | Average 2016-2018 (proportion by number), only applies in scenarios where discarding <br> is allowed. |

### 15.6 Reference points

Reference points were defined for this stock at the IBPNeph (ICES, 2015) and no new proposals were made by WKMS YRef4 (ICES, 2016a; 2016b).

Based on the fact that somebiological parameters are poorly known; inconsistent biological sampling; uncertainties about the stability of the stock over the reference period and uncertainties about the variability of recruitment it is expected that a combined sex $\mathrm{F}_{0.1}$ is a suitable $\mathrm{F}_{\text {MSY }}$ proxy for this stock. This corresponds to a harvest rate of $11 \%$ and this value is expected to deliver high long-term yield with a low probability of recruitment over-fishing. These calculations assume that the UWTV survey has knife-edge selectivity at 17 mm and that the supplied length frequencies represented the population in equilibrium. Currently this fishery is being harvested at $3.14 \%$ (Fsq_2016-2018 $=3.14 \%$; F2018 $=2.55 \%$ ), and historically the available data show a maximum harvest rate of $8.2 \%$ in 2008 which is below the Fmsy proxy.

At the IBP, a MSY Btrigger was defined for this stock. Accordingly, with this definition Btrigger it was set for FU14 as 350 million, corresponded to the abundance observed in 2009.

| Framework | Reference <br> point | Value | Technical basis | Source |
| :--- | :--- | :--- | :--- | :--- |
| MSY approach | MSY $_{\text {trigger }}$ 350 millionindividu- <br> als The lowest observed abundance estimate from the UWTV <br> survey time-series. | ICES (2015) |  |  |
|  | $\mathrm{F}_{\text {MSY }}$ | $11 \%$ harvest rate | $\mathrm{F}_{\text {MSY }}$ proxy equivalent to $\mathrm{F}_{0.1}$ for combined sexes. | ICES (2015) |

### 15.7 Management strategies

There are no explicit management strategies for this stock.

### 15.8 Quality of assessment and forecast

The quality of landings data has improved in the last four years, but concerns over the accuracy of earlier years limits the period we can be confident about regarding trends in LPUE and landings.

Underwater TV surveys have been conducted annually for this stock since 2007. The quality of the data from the first survey and the limited number of valid stations in the survey limits the number of useable surveys to 2008-2019.
The revised algorithm used to derive distance covered by the sledge is considered as significantly more robust than the previous algorithm.

The IBP 2015 managed to address key points:

- Revisions to the area of the Nephrops grounds based on new available data: VMS, UWTV data and sediment information;
- A review of fishery data and raising procedures;
- Review of Reference points: FMSY proxies and MSY Btrigger.

After this revision, the quality of the assessment improved. Although there are still specific uncertainties and assumptions that need to be examined further for the East Irish Sea before less conservative Fmsy proxies could be considered.
There are several key uncertainties and bias sources in the method proposed (these are discussed further in ICES, 2009a). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (ICES, 2007; ICES, 2008; ICES, 2009b). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that is more accurate but no more precise (ICES, 2009a).

The cumulative absolute conversion factor estimates for FU14 are largely based on expert opinion. However, these were based on experience on other grounds and relatively limited experience on these grounds which would make this less reliable. The precision of these cannot yet be characterised. Ultimately, there still remains a degree of subjectivity in the production of UWTV abundance estimates.

The effect of this assumption on realised harvest rates has not been investigated, but remains a key uncertainty.

### 15.9 Recommendation for next benchmark

This stock was last benchmarked by IBPNeph (ICES, 2015). WGCSE will keep the stock under close review and recommend future benchmark as required.

At IBP 2015, it was mentioned that there are specific uncertainties and assumptions that need to be examined further for the East Irish Sea before less conservative $\mathrm{F}_{\text {MSY }}$ proxies could be considered.

- More accurate mapping of the spatial extent of the grounds and fisheries, this includes having positional data for $<12$ metre vessels and more survey data in Wigtown Bay area to better define this ground. Station grid was extended to Wigtown Bay in 2016.
- For now, the total abundance estimate for FU14 is based on the abundance estimates of the geospatial model for the main ground plus adding the area of Wigtown Bay. As this area is becoming a more significant fishing patch, it is worth to consider the use of a separate geospatial model in this ground. This should beexplored in a futurebenchmark work.
- Improvement of spatial coverage and sampling of landings and discards, this includes increasing the sampling levels to cover Northern Irish vessels, as the current sampling is mainly focused on local vessels form Whitehaven port.
- Area specific length-weight and maturity data to validate the parameters used for this FU.
- Better knowledge of the difference in growth and population structure across the area.
- If following the current advice, the recommended catches are taken, then the stock may decrease to well below MSY Btrigger in the short term. The basis for setting MSY Btrigger is currently from recent history may be too high, it could also be due to recent low recruitment (transitory issue) or that the $\mathrm{F}_{\text {MSY }}$ is too high. As such, the MS Ytrigger reference point needs to be looked into. It was noted that the basis for MSY $B_{\text {trigger }}$ was the recent history and that the value may be too high.
- Advice is compiled for ADGNEPH in October. Lagged (one year) TV survey gives good correlation with LPUE, could this be used to calculateharvest rate rather than the in-year ratio?


### 15.10 Management considerations

ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES division level. Management at the Functional Unit level could allow effort and catch to be controlled in line with the scale of the resource.

There are no explicit recruitment indices.
The UWTV survey data allow for the provision of catch options and also to adopt the MSY approach. The UWTV surveys are conducted annually and a benchmark process has been adopted in 2015. In the past, this stock has only been assessed biannually. These data provide the opportunity to reassess this stock more reliably on an annual basis.

### 15.11 References

EU. 2017. Council Regulation (EU) 2017/127 of 20 January 2017 fixing for 2017 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters. Available at [https://eur-lex.europa.eu/legal-con-tent/EN/TXT/?uri=CELEX:02017R0127-20170101](https://eur-lex.europa.eu/legal-con-tent/EN/TXT/?uri=CELEX:02017R0127-20170101) [Accessed: 25/09/2018].

EU. 2018. Council Regulation (EU) 2018/120 of 23 January 2018 fixing for 2018 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters, and amending Regulation (EU) 2017/127. Available at [https://eur-lex.eu-ropa.eu/legal-content/EN/TXT/?uri=CELEX\%3A32018R0120](https://eur-lex.eu-ropa.eu/legal-content/EN/TXT/?uri=CELEX%5C%3A32018R0120) [Accessed: 25/09/2018].
ICES. 2007. Report of the Workshop on the use of UWTV surveys for determining abundance in Nephrops stocks throughout European waters (WKNEPHTV). ICES CM: 2007/ACFM:14.
ICES. 2008. Report of the Workshop and training course on Nephrops burrow identification (WKNEPHBID). ICES CM 2008/LRC:03.
ICES 2009a. Report of Benchmark Workshop on Nephrops assessment (WKNEPH). ICES CM:2009/ACOM:33.
ICES. 2009b. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 2009/LRC: 15, pp 52.
ICES. 2012. Report of the Study Group on Nephrops Surveys (SGNEPS), 6-8 March 2012, Acona, Italy. ICES CM 2012/SSGESST:19. 36 pp.
ICES. 2014. Report of the Working Group on Nephrops Surveys (WGNEPS). ICES CM 2014/SSGESST:20. 57 pp .
ICES. 2015. Report of the Inter-Benchmark Protocol of Nephrops in FU 17 and 14 (IBPNeph), from June to September 2015, by correspondence. ICES CM 2015/ACOM:38. 86 pp.
ICES. 2016a. EU request to ICES to provide FMSY ranges for selected stocks in ICES Subareas 5 to 10. In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.2.3.1.
ICES. 2016b. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

Table 3.8.1. Irish Sea: Landings (tonnes) by FU, 2000-2012. 2017* refers to preliminary landings data. In 2012 and 2013 landings outside FU for Area 7a were not provided, so have been calculated from ICES official landings for 7a minus the FU areas.

| YEAR | FU14 | FU15 | OTHER | TOTAL |
| :---: | :---: | :---: | :---: | :---: |
| 2000 | 567 | 8370 | 1 | 8938 |
| 2001 | 532 | 7441 | 3 | 7976 |
| 2002 | 577 | 6793 | 1 | 7371 |
| 2003 | 376 | 7052 | 3 | 7431 |
| 2004 | 472 | 7267 | 25 | 7764 |
| 2005 | 570 | 6554 | 103 | 7227 |
| 2006 | 628 | 7561 | 52 | 8241 |
| 2007 | 959 | 8491 | 83 | 9533 |
| 2008 | 676 | 1050 | 122 | 11306 |
| 2009 | 708 | 9198 | 57 | 9963 |
| 2010 | 582 | 8963 | 23 | 9568 |
| 2011 | 561 | 10162 | 61 | 10784 |
| 2012 | 531 | 10527 | 208 | 11266 |
| 2013 | 495 | 8672 | 89 | 9256 |
| 2014 | 679 | 8613 | NA | 9292 |
| 2015 | 378 | 8632 | NA | 9010 |
| 2016 | 237 | 7327 | 9 | 7564 |
| 2017* | 265 | 6149 | NA | 6414 |

Table 3.8.2. Irish Sea East (FU14): Landings (tonnes) by country and total discards, 2000-2018.

| Year | Rep. Of Ireland | UK | Other Countries | Total Landings | Discards |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 114 | 451 | 2 | 567 | 80 |
| 2001 | 26 | 506 | 0 | 532 | 42 |
| 2002 | 203 | 373 | 1 | 577 | 42 |
| 2003 | 69 | 306 | 1 | 376 | 11 |
| 2004 | 62 | 409 | 1 | 472 | 28 |
| 2005 | 34 | 536 | 0 | 570 | 33 |
| 2006 | 34 | 594 | 0 | 628 | 22 |
| 2007 | 86 | 873 | 0 | 959 | 47 |
| 2008 | 29 | 652 | 0 | 681 | 37 |
| 2009 | 16 | 692 | 0 | 708 | 6 |
| 2010 | 45 | 538 | 0 | 583 | 9 |
| 2011 | 31 | 530 | 0 | 561 | 0 |
| 2012 | 53 | 478 | 0 | 531 | 0 |
| 2013 | 35 | 460 | 0 | 495 | 38 |
| 2014 | 31 | 648 | 0 | 679 | 35 |
| 2015 | 88 | 290 | 0 | 378 | 18 |
| 2016 | 21 | 216 | 0 | 237 | 20 |
| 2017 | 7 | 258 | 0 | 265 | 28 |
| 2018 | 5 | 263 | 0 | 268 | 9 |

Table 3.8.3. Irish Sea East (FU14): Effort data for the UK and Irish trawl Nephrops directed fleet.

|  | UK direct fleet |  |  | Irish direct fleet |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | EFFORT (KW DAYS) | LANDINGS (TONNES) | LPUE | EFFORT (KW DAYS) | LANDINGS (TONNES) | LPUE |
| 2000 | 145,794 | 392,925 | 2.7 | 47,958 | 109,046 | 2.3 |
| 2001 | 141,686 | 417,382 | 2.9 | 8,691 | 21,242 | 2.4 |
| 2002 | 97,368 | 285,106 | 2.9 | 72,588 | 201,108 | 2.8 |
| 2003 | 114,096 | 225,573 | 2.0 | 23,269 | 41,097 | 1.8 |
| 2004 | 107,570 | 322,982 | 3.0 | 26,345 | 54,810 | 2.1 |
| 2005 | 124,349 | 395,041 | 3.2 | 17,504 | 33,975 | 1.9 |
| 2006 | 249,846 | 407,773 | 1.6 | 6,509 | 18,331 | 2.8 |
| 2007 | 345,818 | 668,017 | 1.9 | 25,309 | 79,193 | 3.1 |
| 2008 | 308,427 | 507,761 | 1.6 | 7,785 | 14,888 | 1.9 |
| 2009 | 262,030 | 499,174 | 1.9 | 5,282 | 13,069 | 2.5 |
| 2010 | 217,937 | 356,188 | 1.6 | 13,496 | 44,615 | 3.3 |
| 2011 | 188,876 | 355,672 | 1.9 | 8,181 | 29,734 | 3.6 |
| 2012 | 163,110 | 301,146 | 1.8 | 20,288 | 52,755 | 2.6 |
| 2013 | 170,799 | 339,429 | 2.0 | 11,304 | 35,459 | 3.1 |
| 2014 | 179,356 | 403,720 | 2.3 | 10,259 | 28,507 | 2.8 |
| 2015 | 79,960 | 155,122 | 1.9 | 27,128 | 83,714 | 3.1 |
| 2016 | 59,970 | 100,733 | 1.7 | 9,496 | 21,225 | 2.2 |
| 2017 | 42,461 | 98,292 | 2.3 | 2,620 | 6,710 | 2.6 |
| 2018 | 58,264 | 112,751 | 1.9 | 3,042 | 5,176 | 1.7 |

Table 3.8.4. Irish Sea East (FU14): Mean size (CL) and weight combined by sex for total annual landings and discards and proportion discarded.

| Year | Mean CL (mm) Landings | Mean CL (mm) Discards | Mean Weight (g) Landings | Mean Weight (g) Discards | Proportion discarded |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 29.83 | 22.32 | 19.05 | 7.52 | 0.26 |
| 2001 | 30.59 | 22.74 | 20.87 | 7.97 | 0.17 |
| 2002 | 30.64 | 23.75 | 22.41 | 8.98 | 0.15 |
| 2003 | 33.69 | 22.43 | 29.12 | 7.62 | 0.10 |
| 2004 | 31.01 | 22.24 | 21.93 | 7.57 | 0.15 |
| 2005 | 30.74 | 23.16 | 21.48 | 8.44 | 0.13 |
| 2006 | 32.36 | 22.75 | 25.07 | 7.98 | 0.10 |
| 2007 | 31.81 | 21.92 | 23.94 | 7.33 | 0.14 |
| 2008 | 31.07 | 23.14 | 22.88 | 8.49 | 0.13 |
| 2009 | 35.57 | 23.21 | 36.49 | 8.58 | 0.04 |
| 2010* |  |  |  |  |  |
| 2011* |  |  |  |  |  |
| 2012* |  |  |  |  |  |
| 2013 | 30.14 | 22.43 | 19.94 | 7.87 | 0.16 |
| 2014 | 31.01 | 24.34 | 22.37 | 9.60 | 0.11 |
| 2015 | 32.05 | 22.57 | 25.19 | 7.82 | 0.13 |
| 2016** | 27.39 | 23.11 | 15.82 | 8.38 | 0.14 |
| 2017 | 29.05 | 24.07 | 18.97 | 9.50 | 0.18 |
| 2018 | 30.58 | 24.46 | 21.39 | 9.78 | 0.07 |

* Values for 2010, 2011 and 2012 are not reliable due to poor sampling.
** Values for 2016 revised at WGCSE 2018 due to inclusion of Northern Irish sampling in 2016 and 2017.

Table 3.8.5. Irish Sea East (FU14): Sumary table for forecast inputs (current used shaded in blue) and historical estimates of raised landings and discards, mean weight in landings and harvest rate.

|  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \frac{1}{\pi} \\ & \stackrel{1}{\infty} \end{aligned}$ | millions | millions | millions | \% | \% | millions |  | \% | tonnes | tonnes | gramme | gramme |
| 2000 | 30 | 11 | 40 | 24.4 | 26.4 |  |  |  | 567 | 80 | 19.05 | 7.52 |
| 2001 | 26 | 5 | 31 | 15.5 | 17.0 |  |  |  | 532 | 42 | 20.87 | 7.97 |
| 2002 | 26 | 5 | 30 | 14.1 | 15.4 |  |  |  | 577 | 42 | 22.41 | 8.98 |
| 2003 | 13 | 1 | 14 | 9.0 | 9.9 |  |  |  | 376 | 11 | 29.39 | 7.64 |
| 2004 | 22 | 4 | 25 | 13.5 | 14.8 |  |  |  | 472 | 28 | 21.93 | 7.57 |
| 2005 | 275 | 4 | 30 | 11.8 | 13.0 |  |  |  | 570 | 33 | 21.48 | 8.44 |
| 2006 | 25 | 3 | 28 | 9.2 | 10.1 |  |  |  | 628 | 22 | 25.07 | 7.98 |
| 2007 | 40 | 6 | 46 | 12.5 | 13.8 |  |  |  | 959 | 47 | 23.94 | 7.33 |
| 2008 | 30 | 4 | 34 | 11.6 | 12.7 | 408 | 63 | 8.2 | 676 | 37 | 22.88 | 8.49 |
| 2009 | 19 | 1 | 20 | 3.3 | 3.7 | 350 | 76 | 5.7 | 707 | 6 | 36.49 | 8.58 |
| 2010 |  |  |  |  |  | 422 | 103 |  | 582 |  |  |  |
| 2011 |  |  |  |  |  | 449 | 99 |  | 561 |  |  |  |
| 2012 |  |  |  |  |  | 694 | 99 |  | 531 |  |  |  |
| 2013 | 25 | 5 | 30 | 15.0 | 16.4 | 487 | 82 | 6.0 | 495 | 39 | 19.94 | 7.87 |
| 2014 | 30 | 4 | 34 | 9.8 | 10.8 | 449 | 92 | 7.5 | 679 | 32 | 22.37 | 9.60 |
| 2015 | 15 | 2 | 17 | 11.9 | 13.0 | 591 | 86 | 2.9 | 378 | 18 | 25.19 | 7.82 |
| 2016* | 15 | 2 | 17 | 12.4 | 13.6 | 430 | 106 | 4.0 | 237 | 20 | 15.82 | 8.38 |
| 2017 | 14 | 3 | 17 | 16.2 | 17.6 | 580 | 89 | 2.9 | 265 | 29 | 18.97 | 9.50 |
| 2018 | 12 | 1 | 13 | 6.3 | 6.9 | 399 | 66 | 2.6 | 268 | 9 | 21.39 | 9.78 |

Note: Abundance is adjusted by using a cumulative absolute conversion factor of 1.2. Abundance (millions) including Wigtown Bay (1.9\% 2008-2010; 6.6\% 2011-2018). Due to poor sampling no estimates for 2010-2012.

* Values for 2016 revised at WGCSE 2018 due to inclusion of Northern Irish sampling in 2016 and 2017.

Table 3.8.6. Nephrops, Irish Sea East (FU14): Results of the 2008-2018 TV surveys (values adjusted for bias).

| Year | No valid stations | Mean Kriged density (no./m) | Abundance (millions) including Wigtown Bay (1.9\% 2008-2010) | Abundance (millions) including Wigtown Bay (6.6\% 2011-2018) | $\begin{aligned} & 95 \% \\ & \text { CI } \end{aligned}$ | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 |  |  | Unreliable data |  |  |  |
| 2008 | 32 | 0.38 | 408 |  | 63 |  |
| 2009 | 32 | 0.33 | 350 |  | 76 |  |
| 2010 | 26 | 0.4 | 422 |  | 103 |  |
| 2011 | 26 | 0.41 |  | 449 | 99 | 11.2\% |
| 2012 | 26 | 0.64 |  | 694 | 99 | 7.3\% |
| 2013 | 31 | 0.45 |  | 487 | 82 | 8.5\% |
| 2014 | 34 | 0.41 |  | 449 | 92 | 10.4\% |
| 2015 | 42 | 0.54 |  | 591 | 86 | 7.4\% |
| 2016 | 48 | 0.40 |  | 430 | 106 | 12.6\% |
| 2017 | 45 | 0.53 |  | 580 | 89 | 7.8\% |
| 2018 | 46 | 0.47 |  | 514 | 118 | 11.7\% |
| 2019 | 41 | 0.30 |  | 399 | 69 | 8.8\% |

Note: Abundance is adjusted by using a cumulative absolute conversion factor of 1.2. Abundance (millions) including Wigtown Bay (1.9\% 2008-2010; 6.6\% 2011-2018).


Figure 3.8.1. Irish Sea East (FU14): Landings in tonnes by country. GBE=England; GBN=Northern Ireland; GBS=Scotland; Rep. of Ireland=Republic of Ireland.


Figure 3.8.2. Irish Sea East (FU14): Effort data (KW days) for UK directed Nephrops fleet.

## Length frequencies for catch (dotted) and landed(solid): Nephrops in fu14



Figure 3.8.3. Irish Sea East (FU14): Length distribution of landings (solid lines) and catch (dotted lines), 20002017. Length frequencies for 2010-2012 are based in very poor sampling so not reliable. Figure shows a vertical display of MLS ( 20 mm CL ) and 35 mm CL levels.

FU14 combined year and mesh


Figure 3.8.4. Irish Sea East (FU14): Final discard ogive pooled for all years (2003-2014) and mesh sizes. L50=23.54 and L25=24.77, (IBPNeph 2015).


Figure 3.8.5. Irish Sea East (FU14): Proportion of males in catch since 1999. Between 2010 and 2012 due to poor sampling levels, estimates of sex ratio are not reliable.


Figure 3.8.6. Irish Sea East (FU14): Mean weight (g) combined by sex for total annual landings and discards. Values for 2010, 2011 and 2012 are not reliable due to poor sampling.

FU14 2019 Grid


Figure 3.8.7. Irish Sea East (FU14): UWTV Survey stations for 2019.


Figure 3.8.8. Irish Sea East (FU14): Co-kriging approach. Interpolation result of VMS (cut off 3\%), survey density (2013-2015) data and mud distribution. A - model output; B - final polygon.



Figure 3.8.9. Irish Sea East (FU14): Burrow density estimates from the UWTV Survey 2008-2018 (individuals/ $\mathrm{m}^{2}$.) Abundance estimates (millions) given at the bottom of each plot are adjusted with the cumulative absolute conversion factor (but does not contain the additional area for Wigtown Bay). Area of ground $=\mathbf{1 0 3 2 . 7 5} \mathrm{Km}^{2}$ for 2008-2010 and 1019.79 $\mathrm{Km}^{2}$ for 2011-2019.


Figure 3.8.10. Irish Sea East (FU14): Burrow density estimates from the UWTV Survey 2008-2019. Btrigger $\operatorname{set}$ as 350 million (orange dashed line).


Figure 3.8.11. Irish Sea East (FU14): Harvest Rate (\% dead removed/UWTV abundance). The dashed and solid lines are the MSY proxy ( $11 \%$ ) and the harvest rate respectively. Between 2010 and 2012 due to poor sampling levels, harvest rate estimates are not reliable.

# 16 Norway lobster (Nephrops norvegicus)in Division 7.a, Functional Unit 15 (Irish Sea, west) 

## Type of assessment

The assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the general process defined by WKNEPH (2009) described in the stock annex. The TV survey is due to be repeated in the summer of 2019 and the new survey will form the basis of advice for this stock in the autumn.

## ICES advice applicable to 2018

ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2014-2016, catches in 2018 should be no more than 11807 tonnes.

To ensure that the stock in Functional Unit 15 is exploited sustainably, management should be implemented at the functional unit level.

## ICES advice applicable to 2019

ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2015-2017, catches in 2019 should be no more than 11107 tonnes.

To ensure that the stock in Functional Unit 15 is exploited sustainably, management should be implemented at the functional unit level.

### 16.1 General

## Stock description and management units

The Irish Sea West (FU15) is comprised of ICES rectangles 35E3-E5, 36E3-E5, 37E3-E5 and 38E4 within 7a. It is included in ICES Area 7 together with the Irish Sea East (FU14), Porcupine Bank (FU16), Aran Grounds (FU17) northwest Irish Coast (FU18), southeast and southwest Irish Coast (FU19), NW Labadie, Baltimore and Galley, and Jones and Cockburn (FU20-21) and the Smalls (FU22).

A TAC is in place for ICES Area 7, which does not correspond to the assessment units. As Nephrops are limited to muddy habitats the distribution of suitable sediment defines the species distribution and the stocks are therefore assessed as seven separate Functional Units. The TAC for Area 7 is shown in the tables section.

## Fishery description

TheFU15 Nephrops fishery firstdeveloped in the late1950s. Theenvironment in the Western Irish Sea is very suitable for Nephrops, with a large mud patch and a gyre that retains the larvae over the mud patch, thus ensuring good recruitment. The ground can be characterized as an area of
very high densities of small Nephrops. Northern Ireland and Ireland are the main countries involved in the FU15 Nephrops fishery.

## The fishery in 2017

The Nephrops fishery in the Irish Sea west is economically the most important in ICES Division 7. a and is mainly prosecuted by vessels from UK (Northern Ireland) and Ireland. Working Group landings from FU15 are presented in Table 17.1 and Figure 17.1. Total declared international Nephrops landings reported from FU15 in 2018 was 5756 t , which are the lowest observed landings since 1980 (Table 17.1). There has been a trend for Irish, since 2012, and more recently Northern Irish vessels to switch to multi (quad) rig trawls. Provisional data suggest a $\sim 30 \%$ increase in Nephrops catch rates and a reduction in fish bycatch of $\sim 30 \%$ due to the lower headline height. Since March 2012, it is mandatory for all Irish vessels to use specified species selective gears. Similar conditions have been introduced in October 2012 for the UK (Northern Ireland) vessels. In 2018, there was marked decrease in LPUE in 2018 (Table 17.2). Recent decline in landings is considered to reflect overall decline in fishing effort with reports of a number of vessels leaving the fishery.

Further general information on the fishery can be found in the stock annex.

## Information from stakeholders

No information from stakeholders.

### 16.2 Data

Commercial size composition data for landings and discards were provided by Northern Ireland and Ireland. Other biological data used in the assessment were as listed in the stock annex compiled by the Benchmark meeting WKNEPH (2009).

## InterCatch

Data were available in InterCatch and used to derive assessment input data.

## Landings

Working Group landings from FU15 are presented in Table 17.1 and Figure 19.1. Total declared international Nephrops landings reported from FU15 in 2018 was 5756 t, which are the lowest observed landings since 1980. Ireland's landings were 1387 t , a small increase from 2017. UK vessels landed 4369 tin 2018, a decrease of $11 \%$ compared tolandings in 2017, landings by Northern Irish vessel contributed to over $98 \%$ of these landings.

## Effort

Effort by the UK fleet remained relatively stable since 2002 following a steady decline from the early 1990s. There was a further marginal reduction in effort and LPUE time-series for Ireland (Table 17.3) compared to 2016, with effort at the lowest reported value in the series. In previous years, these interannual fluctuations have been attributed to the high mobility and flexibility, in terms of fishing in other areas within the TAC area, whereas the Northern Irish effort is mostly concentrated on FU15. Fishing activity from the Irish fleet in FU15 increasingly concentrates on good fishing periods during the year, resulting in a larger and increasing LPUE. The LPUE and
effort LPUE series for Northern Ireland are updated to provide kW days ( kWd ) and LPUE as $\mathrm{kg} / \mathrm{kWd}$. A change to e-logbooks and recording of fishing hours after 2013 means that the recent data are not comparable with the historic series. Recent LPUE and effort after 2013 has remained stable. The LPUE for the Northern Irish and Irish fleets in 2018 were similar $2.10 \mathrm{~kg} / \mathrm{kWd}$ and $2.30 \mathrm{~kg} / \mathrm{kWd}$ but both declines since 2017 from $3.05 \mathrm{~kg} / \mathrm{kWd}$ and $2.43 \mathrm{~kg} / \mathrm{kWd}$.

## Sampling levels

Sampling catches by means of the fisher self-sampling scheme for Northern Irish vessels has continued at sustained high levels with 107 samples collected from the reference fleet, with 34, 25,29 and 19 samples in quarters $1-4$ respectively. The number of discard and catch samples collected from the Irish fleet was seven, six, nine and four samples collected in quarters 1-4 respectively. These rates correspond to one sample per 40.1 t landed by the Northern Irish fleet and one sample for every 53.3 t landed by the Irish fleet. Sampling levels compared to previous years are lower as a number of vessel in the fisher self-sampling scheme left the fleet.

## Commercial length-frequency distributions

Length and sex compositions of Nephrops landed from the Irish Sea West areestimated from port sampling by Ireland and Northern Ireland. Sampling of Northern Ireland catches was not possible during 2003-2007, with the Irish length frequencies raised to the international catch for these years. Northern Ireland sampling resumed in 2008 and these data are combined with those from Ireland for that year.

This Northern Irish fisher self-sampling scheme uses a reference fleet of vessels selected vessels from the main Northern Irish ports. The reference vessels selection is designed to be representative of the entire fleet with systematic rota sampling. The mean sizes of Nephrops in the catches of both the Northern Ireland and Ireland fisheries have fluctuated for the last decade (Tables 19.4-19.5; Figure 17.1). There is little evidence to suggest a long-term trend in the mean size of males and females in the landings and catches, which continues to fluctuate around the series mean (Figure 17.2).

## Sex ratio

The sex ratioby year is shown in Figure 17.3. This shows some fluctuations over time. In general, the sex ratio in landings and catches are biased toward males, with a geomean of $56.2 \%$ males in landings (1986-2018) and $52.4 \%$ in catches (1986-2018). A bias toward males in catches was observed in 2017 comprising $61 \%$ in landings and $55.6 \%$ in the catch compared to $64.3 \%$ in the catch and $59.2 \%$ in landings in 2018. The stronger bias of males in landings relates to the average larger size of male Nephrops.

## Mean weights

Explorations of the mean weight in the catch samples by sex shows a strong seasonal pattern in the females (Figure 17.4). This corresponds with the emergence of mature females from the burrows to mate in summer. There is no evidence of a recent trend toward decreasing mean weights (Figure 17.5), however compared to the early part of the time-series mean weights have decreased. The mean weights in landings (2016-2018) and mean weights in discards (2016-2018) are used in the basis for calculating catch options (Section 19.4).

## Discards

Annual discard rates are estimated using unsorted catch and discards sampling. Unsorted catches and samples of retained catch are provided by vessels. The catch sample is partitioned into landings and discards using a discard selection ogive. This selection ogive can be derived per sample or as aggregation of samples within a quarter or year when sampling rates are low. Sampling effort is stratified weekly, but quarterly aggregations are used for quarterly length frequencies and discard estimates. The length-weight regression parameters given in the stock annex are used to calculate sampled weights and appropriate raising factors. Discarding practice is highly variable, mainly driven by market demand, and was $29 \%$ of the catch by number in 2018(Table 17.6). A discard survival rate of $10 \%$ is assumed for Nephropsfrom this FU (WKNEPH 2009).

## Surveys

## Abundance indices from UWTV surveys

Since 2003, Ireland and Northern Ireland have jointly carried out underwater television surveys of the main Nephrops grounds in the western Irish Sea. These surveys were based on a randomised fixed-grid design. The methods used during the surveys were similar to those employed for UWTV surveys of other Nephrops stocks and were as agreed by WKNEPHTV (ICES, 2007), WKNEPBID (ICES, 2008), SGNEPS (ICES, 2009; 2010; 2012), WKNEPH (ICES, 2009) and WGNEPS (ICES, 2013; 2014; 2015; 2016). From 2003 to 2011 year an average of 146 valid stations was covered by the two surveys combined and the data were raised to a stock area of around $5290 \times 10-6 \mathrm{~km}^{2}$ as detailed in Table 17.7. Details of the survey methodology are available in WGNEPS (ICES, 2016). Figure 17.6 shows the distribution of stations sampled in 2019. The number of stations were significantly reduced in 2012 following a recommendation from SGNEPS 2012 that a CV (or relative standard error) of $<20 \%$ is an acceptable precision level for UWTV survey estimates of abundance. This allowed sampling intensity tobe reduced, and survey effort allocated to other areas and FUs in area 7. Figures 17.7-17.9 are contour plot of the kriggeddensity estimates for FU15 over the period 2003-2019. The resulting krigged burrow abundance estimate was 4.4 billion burrows. This was a similar result of that obtained in 2015, but a $10 \%$ lower than the abundance in 2018. In contrast to 2017, the spatial distribution of burrows shows a high density band on the central western area of the survey ground. (Table 17.7). A violin plot of the burrow densities observed in the survey (2003-2019) is shown in Figure 17.10. The character of the burrow densities encountered has remained consistent over time; characterised by a relatively high occurrence of low density stations and a normal distribution densities around one burrow $/ \mathrm{m}^{2}$. Confidence in the survey estimates and design are assured through the maintained low coefficient of variation on the burrow estimates.

The use of the UWTV surveys for the provision of Nephrops management advice was extensively reviewed by WKNEPH (ICES, 2009) and potential biases were highlighted including those due to edge effects; species burrow misidentification and burrow occupancy. A cumulative bias correction factor estimated for FU15 was 1.14 , which means the TV survey is likely to overestimate Nephrops abundance by $14 \%$.

## Nephrops trawl surveys

In addition to UWTV surveys, Northern Ireland have completed spring (April) and summer (August) Nephrops trawl surveys since 1994 and provide data on catch rates, size composition and biological data from fixed stations in the western Irish Sea as detailed in the Stock Annex (Stock Annex Figure 1). Survey CPUE has remained stable over time. Mean carapace length-by-
sex (from the trawl survey) shows inter-annual variation fluctuating around mean with no apparent trend over time (Figure 17.11).

Due to reduced resources, the spring survey series was terminated in 2010 as part of a national rationalisation of the survey programme after considering benefits to management and stock assessment. Due to a major ship breakdown, no data are available for the 2013 summer survey. The summer trawl survey catch rates correlate somewhat with UWTV survey abundance estimates (Figure 17.12), but showed a deviating trend, especially in 2010. The longer time-series of the trawl survey shows that catch rates in the last few years $(2005-2009,2011)$ are close to the mean of the series when UWTV burrow abundances were in the range of 5-6 billion burrows. The reduction in the 2010 trawl estimate, that showed a conflicting trend to the UWTV abundance, is most likely associated with the survey taking place in suboptimal tidal conditions. Usually the trawl survey coincides with slack tides, but this was not optimal in 2010 due to availability of the ship and synchronisation with the UWTV survey.

### 16.3 Assessment

## Comparison with previous assessments

The assessment approach used by WGCSE 2018 is consistent with that set out in the stock annex and WKNEPH (WKNEPH, 2009). Since the most recent three years of sampling data were available, three-year averages of mean weights in the landings and proportions retained in the fishery have been used. This is in line with the procedure used for other stocks in areas 6 and 7 by WGCSE.

## State of the stock

The stock size is estimated to show a decrease, but within the limits previously observed for the stock. The harvest ratio has decreased further in 2018 and is below FmsY (Figure 17.13). This stock has previously sustained landings at around $9000 t$ for many years. The stock increased until 2003, with a general decrease until 2014 and has increased since then. The most recent UWTV abundance estimate of 4.4 billion in 2018 follows a period (2016-2017) of above average size. The geometric mean of current series is 4.92 billion. Figure 17.13 is the stock summary plot for FU15. Recent harvest rates have fluctuated around Fmsy, but is estimated as 10.0 in 2018, having decreased from 19.9 in 2015 (Table 17.6). The stock is estimated to be well above $B_{\text {trigger }}$ ( 3000 million).

### 16.4 Catch option table

Catch option table inputs are presented in Table 17.6 and summarised below. A three-year average (2016-2018) of mean weight in the landings and proportion of removals retained was used.
A landings prediction for 2020 was made for FU15 using the approach agreed at the Benchmark Workshop (WKNEPH, 2009) and outlined in the stock annex made on the basis of the 2019 UWTV survey.

The basis for the catch options.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Stock abundance | 4.4 billion indi- <br> viduals | UWTV survey 2019. |
| Mean weight in <br> landings | 14.5 g | Average 2016-2018. |
| Mean weight in <br> discards | 7.8 g | Average 2016-2018. |
| Discard rate | Average 2016-2018 (by number).Calculated as discards divided by landings + <br> discards. |  |
| Discard survival <br> rate | Only applies in scenarios where discarding is allowed. |  |
| Dead discard rate | Average 2016-2018 (by number). Calculated as dead discards divided by dead <br> removals (landings + dead discards). Only applies in scenarios where discarding <br> is allowed. |  |

### 16.5 Reference points

A decision-making framework for the choice of FMSY proxy reference points is available in the introduction to the NephropsICES advice sheets. The current FMSY proxy reference points for FU15 Nephrops was evaluated at WKMS YRef4. The MSY reference point for FU15 Nephrops is the Fmax for combined sexes. No precautionary reference points have been defined for Nephrops stocks. Whereas the Fmsy proxy reference points were chosen with the intent that they should lead to a low probability of stock overfishing.

Previously the CPUE data from the trawl surveys were scaled to the UWTV index to provide a $B_{\text {trigger }}$ approximation based on the mean of the five lowest survey catch rates in the time-series (Figure 17.8), this is still accepted as an appropriate Btrigger for FU15.

| Stock code | MSY Flower | F MSY | MSY Fupper with AR | MSY $B_{\text {trigger }}$ | MSY Fupper with no AR |
| :--- | :---: | :---: | :---: | :---: | :---: |
| nep-15 | 12.4 | 18.2 | 18.2 | $3000^{*}$ | 18.2 |

*Abundance in millions.

### 16.6 Management strategy

As yet there are no explicit management strategies for this stock.

### 16.7 Quality of assessment and forecast

Uncertainties in the survey, mean weight in the landings and discard rates are not taken into account in the deterministic catch option. There is some variability in these over time.
There are several key uncertainties and bias sources in the method used here (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007; WKNEPHBID 2008;SGNEPS 2009). These have led to a revision in
the historical time-series of survey abundance estimates for FU15, which was presented to last year's Working Group. Ultimately there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996).

Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that was more accurate but no more precise (WKNEPH 2009). The survey estimates themselves are very precisely estimated (CVs 2-5\%) given the homogeneous distribution of burrow density and the modelling of spatial structuring. The cumulative bias estimates for FU15 are largely based on expert opinion (see Stock Annex). The precision of these bias corrections cannot yet be characterised but is likely to be higher than that observed in the survey.

In the provision of catch options based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. These parameters are quite variable, in future years the uncertainty in these key parameters should be estimated.
The quality of landings data has improved since 2007 with the implementation of sales notes and buyers and sellers legislation. Prior to that, there were concerns that landings were underreported. The harvest ratio may be under estimated prior to 2007.

### 16.8 Recommendations for next benchmark

WGCSE will keep the stock under close review and recommend future benchmark as required.

### 16.9 Management considerations

The FU15 Nephrops fishery first developed in the late 1950s. Since then it has sustained landings of around 8500 t for more than 30 years. Fishing effort in the past has been very high but has declined somewhat in recent years. The environment in the western Irish Sea is very suitable for Nephrops with a large mud patch and gyre, which retains the larvae over the mud patch thus ensuring good recruitment. The ground can be characterised as an area of very high densities of small Nephrops. All available information indicates that size structure of catches appears to have changed little since the fishery first began.

The Nephrops trawl fisheries take bycatches of other species, especially juvenile whiting, but also cod. Catches of these species should be reduced to as low as possible because of the poor status of these stocks. A conditional national licence has been introduced by Ireland since March 2012, making the use of grids or separator panels mandatory for all TR2 boats fishing in the Irish Sea. Around $55 \%$ of the Irish vessels use separator trawls and while $45 \%$ have opted to use Swedish grids to reduce bycatch. Additionally, there has been a trend for Irish vessels to switch to multi (quad) rig trawls. Provisional data suggests a $\sim 30 \%$ increase in Nephrops catch rates and a reduction in fish bycatch of $\sim 30 \%$ due to the lower headline height.

Since October 2012, all TR2 vessels in the UK (Northern Ireland) fleet are required to use a highly selective fishing gear. In the Irish Sea, these currently include Seltra 300 mm box trawl, 270 mm diamond mesh panel Seltra box trawl and 300 mm square mesh panel. All these gears are being developed with the aim of achieving exemption from the cod recovery plan under Article 11 (less than $1.5 \%$ cod catch). Enforcement is through the issue cod recovery zone fishing authorisations, where no authorisation is given to a vessel that is not using a highly selective gear.
ICES has repeatedly advised that managementshould be at a smaller scale than the ICES Subarea 7. Management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort are at the same scale as the resource.

A number of cod recovery measures have been introduced since 2000 to promote recovery of IrishSea cod stocks. These include a closure of the western IrishSea cod spawning grounds from mid-February to end of April since 2000, with a later extension to the eastern Irish Sea closure. Despite a partial derogation for Nephrops vessels during the closed period, the distribution of effort on Nephrops has been affected by this management plan. There have also been decommissioning schemes to reduce fishing effort.

### 16.10 References

Briggs, R.P., Armstrong, M.J., Dickey-Collas, M., Allen, McQuaid, N. and Whitmore, J. 2002. Estimation of Nephrops Biomass in the Western Irish Sea from Annual Larval Production. ICES Journal of Marine Research, 59: 109-119.

Hill, A. E., Durazo, R. and Smeed, D. A. 1994. Observations of a cyclonic gyre in the western Irish Sea. Continental Shelf Research, 14: 479-490.

Hill, A.E., Brown, J. and Fernand, L. 1996. The western Irish Sea gyre: a retention mechanism for the Norway Lobster (Nephrops norvegicus)? Oceanologica Acta 19:357-369.

McQuaid, N., Briggs R. P. and Roberts D. 2009. Fecundity of Nephrops norvegicus from the Irish Sea. Journal of the Marine Biological Association of the UK, 89: 1181-1188.

Marrs, S.J., Atkinson, R.J.A., Smith, C.J. and Hills, J.M. 1996. Calibration of the towed underwater TV technique for use in stock assessment of Nephropsnorvegicus. Reference no. 94/069 (Study Project in support of the Common Fisheries Policy XIV/1810/C1/94, call for proposals 94/C 144/04).

ICES. 2016XX. EU request to ICES to provide FMSY ranges for selected stocks in ICES subareas 5 to 10. In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.2.3.1.

ICES. 2016YY. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.
SGNEPS. 2009. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 2009/LRC.
WKNEPH. 2009. Report of the Benchmark Workshop on Nephrops assessment (WKNEPH). ICES CM 2009/ACOM:33.

WKNEPHBID. 2008. Report of the Workshop and training course on Nephrops burrow identification (WKNEPHBID). ICES CM 2008/LRC:03.

WKNEPHTV. 2007. Report of the Workshop on the use of UWTV surveys for determining abundance in Nephrops stocks throughout European waters (WKNEPHTV). ICES CM: 2007/ACFM: 14.

Table 17.1. Irish Sea West (FU15): Landings (tonnes) by country, 2000-2018.

| Year | Ireland | Isle of Man | UK | Other countries | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 3433 | 0 | 4937 | 0 | 8370 |
| 2001 | 2689 | 3 | 4749 | 0 | 7441 |
| 2002 | 2291 | 1 | 4501 | 0 | 6793 |
| 2003 | 2709 | 4 | 4352 | 0 | 7065 |
| 2004 | 2786 | 13 | 4470 | 1 | 7270 |
| 2005 | 2133 | 0 | 4420 | 0 | 6554 |
| 2006 | 2051 | 1 | 5508 | 1 | 7561 |
| 2007 | 2767 | 0 | 5724 | 0 | 8491 |
| 2008 | 3132 | 50 | 7323 | 2 | 10508 |
| 2009 | 2343 | 1 | 6855 | 0 | 9198 |
| 2010 | 2578 | 0 | 6384 | 0 | 8963 |
| 2011 | 3575 | 2 | 6584 | 0 | 10162 |
| 2012 | 3794 | 3 | 6732 | 0.2 | 10529 |
| 2013 | 2465 | 31 | 6175 | 0.2 | 8672 |
| 2014 | 2938 | 0** | 5676 | 0.0 | 8613 |
| 2015 | 2199 | 0** | 6433 | 0.3 | 8632 |
| 2016 | 1609 | 0** | 5715 | 3 | 7327 |
| 2017 | 1253 | 0** | 4896 | 0 | 6150 |
| 2018 | 1387 |  | 4369 | 0 | 5756 |

[^8]Table 17.2. Irish Sea West (FU15): Catches and landings (tonnes), effort (' 000 hours trawling), CPUE and LPUE ( $\mathrm{kg} /$ hour trawling) Republic of Ireland Nephrops Directed Trawlers 2000-2013. Time-series updated in 2018.

| Year | Landings ( Kg ) | Effort (Hours) | Effort (days) | Effort (kwdays) | LPUE |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 1706969 | 44459 | 3516 | 835977 | 2.041885 |
| 1996 | 1406140 | 31409 | 2326 | 607785 | 2.313549 |
| 1997 | 2801501 | 60502 | 4518 | 1124379 | 2.491599 |
| 1998 | 2696979 | 52277 | 4051 | 1053491 | 2.560039 |
| 1999 | 4031508 | 73786 | 5260 | 1367903 | 2.947217 |
| 2000 | 3227565 | 61936 | 4396 | 1199896 | 2.68987 |
| 2001 | 2428587 | 51111 | 3435 | 939387 | 2.585289 |
| 2002 | 2015965 | 46072 | 2900 | 873563 | 2.307749 |
| 2003 | 1620391 | 47704 | 3120 | 878568 | 1.844355 |
| 2004 | 2586760 | 52673 | 3500 | 1033073 | 2.503946 |
| 2005 | 2111185 | 50825 | 3414 | 1003901 | 2.102981 |
| 2006 | 2031881 | 53461 | 3535 | 1084251 | 1.873995 |
| 2007 | 2728841 | 52550 | 3575 | 1056291 | 2.583419 |
| 2008 | 3165781 | 49218 | 3401 | 1027919 | 3.079796 |
| 2009 | 2333433 | 34651 | 2368 | 706178 | 3.304312 |
| 2010 | 2505061 | 36504 | 2546 | 739345 | 3.388218 |
| 2011 | 3554343 | 47640 | 3229 | 921298 | 3.857972 |
| 2012 | 3725318 | 49313 | 3560 | 966006 | 3.856413 |
| 2013 | 2269336 | 33818 | 2571 | 682793 | 3.323608 |
| 2014 | 2449612 | 40371 | 3007 | 852740 | 2.872635 |
| 2015 | 2119880 | 35898 | 2733 | 756719 | 2.80141 |
| 2016 | 1529418 | 28249 | 2301 | 556452 | 2.748516 |
| 2017 | 1120690 | 22516 | 1749 | 410628 | 2.729208 |
| 2018 | 1363910 | 27084 | 1919 | 535002 | 2.549353 |

Table 17.3. Irish Sea West (FU15): Landings (tonnes), effort ('000 hours trawling), LPUE (kg/hour trawling), effort ( ${ }^{\circ} 000 \mathrm{~kW}$ days) and LPUE ( $\mathrm{kg} / \mathrm{kWd}$ ) of Northern Ireland Nephrops trawlers, 2000-2018.

| Year | Landings | Effort ('000 hours) | LPUE ('000 hrs) | kW days ('000) | LPUE kWd |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 4758 | 168.7 | 28.2 |  |  |
| 2001 | 4587 | 163.7 | 28.0 |  |  |
| 2002 | 4495 | 130.8 | 34.4 |  |  |
| 2003 | 4146 | 136.1 | 29.0 |  |  |
| 2004 | 4273 | 144.3 | 29.6 |  |  |
| 2005 | 4235 | 138.4 | 30.6 |  |  |
| 2006 | 5356 | 144.1 | 37.2 |  |  |
| 2007 | 5512 | 126.9 | 43.4 |  |  |
| 2008 | 7056 | 141.4 | 49.9 |  |  |
| 2009 | 6487 | 134.7 | 48.2 |  |  |
| 2010 | 5888 | 141.1 | 41.7 |  |  |
| 2011 | 5952 | 132.7 | 44.9 |  |  |
| 2012 | 5865 | 137.8 | 42.6 |  |  |
| 2013 | 5605 | 135.7 | 41.3 | 2151.9 | 2.60 |
| 2014 | 5190 | 114.6 | 45.3 | 2111.2 | 2.46 |
| 2015 | 6396 |  |  | 1962.6 | 3.26 |
| 2016 | 5638 |  |  | 2107.3 | 2.68 |
| 2017 | 4789 |  |  | 1904.3 | 2.51 |
| 2018* | 4293 |  |  | 2079.3 | 2.06 |

[^9]Table 17.4. Irish Sea West (FU15): Mean sizes (mm CL) of male and female Nephrops in Northern Ireland catches, landings and discards, 2000-2018.

| Year | Catches <br> Males | Females | Landings <br> Males | Females | Discards <br> Males | Females |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |
| 2000 | 27.7 | 24.5 | 29.4 | 26.3 | 22.5 | 22.6 |
| 2001 | 25.7 | 23.6 | 26.1 | 24.4 | 21.7 | 21.2 |
| 2002 | 26.7 | 24.1 | 26.7 | 24.9 | 21.8 | 21.7 |
| 2003 | na | na | na | na | na | na |
| 2004 | na | na | na | na | na | na |
| 2005 | na | na | na | na | na | na |
| 2006 | na | na | na | na | na | na |
| 2007 | na | na | na | na | na | na |
| 2008 | 25.9 | 24.6 | 26.9 | 25.5 | 21.4 | 21.5 |
| 2009 | 27.7 | 25.1 | 29.3 | 26.5 | 23.6 | 23.2 |
| 2010 | 28.3 | 25.6 | 29.5 | 26.3 | 23.2 | 22.8 |
| 2011 | 27.6 | 26.0 | 29.3 | 27.7 | 22.6 | 22.8 |
| 2012 | 26.8 | 24.3 | 27.7 | 25.4 | 21.7 | 21.1 |
| 2013 | 26.2 | 24.2 | 27.2 | 25.4 | 21.5 | 21.3 |
| 2014 | 26.3 | 23.9 | 27.1 | 24.9 | 21.1 | 20.6 |
| 2015 | 25.3 | 23.4 | 26.8 | 24.7 | 21.6 | 21.3 |
| 2016 | 25.9 | 24.3 | 26.9 | 25.5 | 22.3 | 21.8 |
| 2017 | 27.0 | 24.8 | 28.0 | 26.1 | 22.9 | 22.5 |
| 2018 | 27.6 | 25.1 | 28.8 | 26.6 | 23.3 | 22.5 |

na $=$ not available.

Table 17.5. Irish Sea West (FU15): Mean sizes (mm CL) of male and female Nephrops in Republic of Ireland catches, landings and discards, 2000-2018.

| Year | Catches |  | Landings |  | Discards |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Males | Females | Males | Females | Males | Females |
| 2000 | 29.1 | 27.1 | 32.2 | 29.7 | 24.3 | 24.0 |
| 2001 | 26.7 | 24.8 | 28.6 | 27.0 | 23.0 | 22.2 |
| 2002 | 28.9 | 25.4 | 30.2 | 27.8 | 24.6 | 23.6 |
| 2003 | 27.7 | 24.9 | 29.7 | 26.9 | 24.0 | 23.1 |
| 2004 | 28.1 | 26.1 | 29.7 | 27.8 | 23.9 | 23.7 |
| 2005 | 28.5 | 26.8 | 30.1 | 29.1 | 23.9 | 23.2 |
| 2006 | 27.7 | 25.5 | 29.5 | 27.1 | 23.8 | 23.1 |
| 2007 | 27.7 | 25.4 | 29.8 | 27.9 | 24.0 | 23.3 |
| 2008 | 27.4 | 24.6 | 28.9 | 26.6 | 22.0 | 21.4 |
| 2009 | 28.5 | 26.3 | 30.5 | 29.2 | 24.3 | 23.4 |
| 2010 | 28.0 | 25.9 | 29.6 | 27.6 | 23.8 | 23.3 |
| 2011 | 27.0 | 25.7 | 28.8 | 27.3 | 23.7 | 23.5 |
| 2012 | 26.8 | 25.6 | 28.3 | 27.0 | 23.2 | 23.0 |
| 2013 | 26.3 | 25.1 | 27.4 | 26.5 | 23.1 | 22.6 |
| 2014 | 27.7 | 24.9 | 29.2 | 26.3 | 23.6 | 23.3 |
| 2015 | 27.7 | 25.7 | 29.5 | 27.4 | 24.4 | 24.0 |
| 2016 | 26.0 | 25.0 | 27.3 | 26.4 | 23.5 | 23.3 |
| 2017 | 27.2 | 25.0 | 28.1 | 26.2 | 23.4 | 22.6 |
| 2018 | 27.4 | 24.9 | 29.8 | 22.8 | 24.6 | 22.8 |

Table 17.6. Irish Sea West (FU15): Proportion discarded by weight and number from FU15. (Note a 10\% survivorship of discards is assumed in HR and forecast calculations).

| Year | Landings in number (millions) | Total discards in number (millions) | Removals in number (millions) | UWTV abundance estimates (billions) | 95\% conf. intervals | Harvest rate | Mean weight in landings (g) | Mean weight in discards (g) | Discard rate by number | Dead discard rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 404 | 291 | 666 | 5.5 | 0.27 | 12.1 | 17.5 | 9.1 | 42\% | 39\% |
| 2004 | 416 | 218 | 612 | 5.5 | 0.3 | 11 | 17.5 | 9.1 | 34\% | 32\% |
| 2005 | 346 | 157 | 488 | 5.7 | 0.44 | 8.6 | 18.9 | 9 | 31\% | 29\% |
| 2006 | 467 | 261 | 701 | 5.4 | 0.41 | 13 | 16.1 | 8.8 | 36\% | 33\% |
| 2007 | 511 | 375 | 848 | 5.1 | 0.34 | 16.5 | 16.5 | 8.7 | 42\% | 40\% |
| 2008 | 755 | 191 | 927 | 4.3 | 0.25 | 21.6 | 13.9 | 7.4 | 20\% | 19\% |
| 2009 | 567 | 335 | 868 | 4.6 | 0.26 | 18.8 | 16.2 | 8.8 | 37\% | 35\% |
| 2010 | 572 | 180 | 733 | 5 | 0.31 | 14.7 | 15.7 | 8.6 | 24\% | 22\% |
| 2011 | 644 | 332 | 943 | 4.9 | 0.23 | 19.4 | 15.8 | 8.1 | 34\% | 32\% |
| 2012 | 771 | 258 | 1003 | 5.1 | 0.29 | 19.8 | 13.7 | 7.2 | 25\% | 23\% |
| 2013 | 662 | 229 | 867 | 4.3 | 0.27 | 20.1 | 13.1 | 7 | 26\% | 24\% |
| 2014 | 641 | 198 | 819 | 4.6 | 0.25 | 17.8 | 13.4 | 7.2 | 24\% | 22\% |
| 2015 | 620 | 280 | 872 | 4.4 | 0.29 | 19.9 | 13.9 | 8.0 | 31\% | 29\% |
| 2016 | 562 | 245 | 783 | 5.1 | 0.3 | 15.4 | 13.0 | 7.9 | 30\% | 28\% |
| 2017 | 426 | 152 | 563 | 5.3 | 0.3 | 10.6 | 14.4 | 8.0 | 26\% | 24\% |
| 2018 | 360 | 145 | 491 | 4.9 | 0.3 | 10.0 | 16.1 | 7.4 | 29\% | 27\% |
| Average 2016-2018 |  |  |  | 4.4 (2019 UWTV) |  |  | 14.5 | 7.7 | 28.5\% | 26.4\% |

Table 17.7. Irish Sea West (FU15): Results from NI/ROI collaborative UWTV surveys of Nephrops grounds in 2003-2019.

| Ground | Year | Number of stations | Mean Density ( $\mathrm{No} . / \mathrm{M}^{2}$ ) | Domain Area (km²) | Estimate <br> (billions) | CV on Burrow estimate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Western Irish Sea | 2003 | 160 | 0.99 | 5295 | 5.5 | 3\% |
|  | 2004 | 147 | 1.00 | 5310 | 5.5 | 3\% |
|  | 2005 | 141 | 1.02 | 5281 | 5.7 | 4\% |
|  | 2006 | 138 | 0.97 | 5194 | 5.4 | 4\% |
|  | 2007 | 148 | 0.93 | 5285 | 5.1 | 3\% |
|  | 2008 | 141 | 0.77 | 5287 | 4.3 | 3\% |
|  | 2009 | 142 | 0.83 | 5267 | 4.6 | 3\% |
|  | 2010 | 149 | 0.90 | 5307 | 5.0 | 3\% |
|  | 2011 | 156 | 0.88 | 5289 | 4.9 | 2\% |
|  | 2012 | 99 | 0.91 | 5291 | 5.1 | 3\% |
|  | 2013 | 80 | 0.78 | 5278 | 4.3 | 3\% |
|  | 2014 | 99 | 0.83 | 5272 | 4.6 | 3\% |
|  | 2015 | 100 | 0.79 | 5279 | 4.4 | 3\% |
|  | 2016 | 100 | 0.84 | 5260 | 5.1 | 3\% |
|  | 2017 | 101 | 0.90 | 5304 | 5.3 | 3\% |
|  | 2018 | 100 | 0.85 | 5791 | 4.9 | 3\% |
|  | 2019 | 100 | 0.76 | 5370 | 4.4 | 3\% |



Figure 17.1. Irish Sea West (FU15): Long-term trends in landings, effort, LPUE, and mean sizes of Nephrops. [The quality of landings data has improved since 2007 with the implementation of sales notes and buyers and sellers legislation, which result in misleading LPUE trend plots pre and post 2007].

## Length frequencies for Landings: <br> Nephrops in FU15_5



Figure 17.2. Irish Sea West (FU15): Length distributions in the landings (solid) and catches (dotted) 1986-2018.


Figure 17.3 Nephrops in FU15 (Irish Sea West). Sex ratio of landings (1986-2018) and catch (19862018).


Figure 17.4 Nephrops in FU15 (Irish Sea West). Mean weight in catch samples by sex with GAM loess smoother ( $k=20$ ).


Figure 17.5 Nephrops in FU15 (Irish Sea West). Mean weight in landings and discards.


Figure 17.6. Irish Sea West (FU15): 2019 UWTV survey stations.


Figure 17.7. Irish Sea West (FU15): Contour plots of the krigged density estimates for the Irish Sea from 2003-2008.


Figure 17.8. Irish Sea West (FU15): Contour plots of the krigged density estimates for the Irish Sea from 2009-2014.


Figure 17.9. Irish Sea West (FU15): Contour plots of the krigged density estimates for the Irish Sea from 2009-2019.


Figure 17.10. Irish Sea West (FU15): Box and kite plot of burrow density observed during UWTV survey 20032019.


Figure 17.11 Irish Sea West (FU15): Nephrops catches (kg per nm) from NI trawl surveys. No data available in 2013 due to ship breakdown.


Figure 17.12. Irish Sea West (FU15): Revised UWTV index and scaled trawl survey. CPUE along with $B_{\text {trigger }}$ based upon mean of five lowest trawl survey values. Abundance figures have not been bias corrected.


Figure 17.13. Norway lobster in Division 7.a, Functional Unit 15. Summary of the stock assessment. Catches (discard data are only available from 1986), harvest rate (sum of landings and dead discards in numbers, divided by total abundance), survey abundance (Underwater TV, billions; SSB proxy; $95 \%$ confidence intervals). Harvest rates between 2003 and 2006 may be unreliable because of underreporting of landings. Orange lines represent MSY $B_{\text {trigger }}$ and the $F_{\text {msy }}$ harvest rate.

# 17 Norway lobster (Nephrops norvegicus) in divisions 7.b-c and 7.j-k, Functional Unit 16 (west and southwest of Ireland, Porcupine Bank) 

## Type of assessment in 2019

Available data on the fishery for 2018 and 2019 and other stock indicators have been updated here according to the stock annex (Nephrops FU16). The assessment and catch options follow the agreed procedures set out in the stock annex.

## ICES advice applicable to 2018

"ICES advises that when the MSY approach is applied, catches in 2018 should be no more than 2734 tonnes."

## ICES advice applicable to 2019

"ICES advises that when the MSY approach is applied, and assuming zerodiscards, catchesin 2019 should be no more than 2645 tonnes.

To ensure that the stock in Functional Unit 16 is exploited sustainably, management should be implemented at the functional unit level."

### 17.1 General

## Stock description and management units

The TAC area is Subarea 7, since 2011 an 'of which' clause was introduced specifically for the Porcupine Bank (FU16) see Table 20.1. The Functional Unit for assessment includes some parts of the following ICES divisions 7.b, c , j , and k . The exact stock area is shown on the map below and includes the following ICES Statistical rectangles:31-35 D5-D6;32-35 D7-D8.


The FU16 outlined by the red line. The closed area from 1 May-31 July since 2010 (reduced to only May since 2013) is shown with a green line. Irish Nephrops directed fishing effort between 2006-2009 derived from integrated VMS and logbook information is shown as a heat map.

## Management applicable to 2018 and 2019

## TAC in 2018

COUNCIL REGULATION (EU) 2018/120 of 23 January 2017 fixing for 2018 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters.


| fo | he limits of the abovementio <br> Functional Unit 16 of ICES <br> Subarea 7 (NEP/*07U16): |
| :---: | :---: |
|  |  |
| Spain | 825 |
| France | 516 |
| Ireland | 992 |
| United Kingdom | 401 |
| Union | 2734 |

## TAC in 2019

COUNCIL REGULATION (EU) 2019/124 of 30 January 2019 fixing for 2019 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters.

| Species: | Norway lobster <br> Nephrops norvegicus |  | Zone: | $\begin{aligned} & 7 \\ & \text { (NEP/07.) } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: |
| Spain |  | $1187{ }^{(1)}$ |  |  |
| France |  | $4811{ }^{(1)}$ |  |  |
| Ireland |  | $7296{ }^{(1)}$ |  |  |
| United Kingdom |  | $6490{ }^{(1)}$ |  |  |
| Union |  | $19784{ }^{(1)}$ |  |  |
| TAC |  | $19784{ }^{(1)}$ |  | Analytical TAC <br> Article 13(1) of this Regulation applies |


| (1)Special condition: within the limits of the abovementioned quotas, no more than the quantities given below may be taken in the <br> following zone: <br> Functional Unit 16 of ICES Subarea <br> $7\left(\mathrm{NEP} /{ }^{*} 07 \mathrm{U} 16\right)$ : |
| :--- |
| Spain |
| France |
| Ireland |
| United Kingdom |
| Union |

### 17.2 Closed area restrictions

A seasonal closed area has been in place for three months May 1-31 July between 2010-2012 (shown in the map above and co-ordinates below). The period of the closure was been reduce to only one month after 2013. Article 11 of COUNCIL REGULATION (EU) 2017/127 is given below:

Article 11<br>Closed fishing seasons

1. It shall be prohibited to fish or retain on board any of the following species in the Porcupine Bank during the period from 1 May to 31 May 2017: cod, megrims, anglerfish, haddock, whiting, hake, Norway lobster, plaice, pollack, saithe, skates and rays, common sole, tusk, blue ling, ling and picked dogfish.

For the purposes of this paragraph, the Porcupine Bank shall comprise the geographical area bounded by rhumb lines sequentially joining the following positions:

| Point | Latitude | Longitude |
| :---: | :---: | :---: |
| 1 | $52^{\circ} 27^{\prime} \mathrm{N}$ | $12^{\circ} 19^{\prime} \mathrm{W}$ |
| 2 | $52^{\circ} 40^{\prime} \mathrm{N}$ | $12^{\circ} 30^{\prime} \mathrm{W}$ |
| 3 | $52^{\circ} 47^{\prime} \mathrm{N}$ | $12^{\circ} 39,600^{\prime} \mathrm{W}$ |
| 4 | $52^{\circ} 47^{\prime} \mathrm{N}$ | $12^{\circ} 56^{\prime} \mathrm{W}$ |
| 5 | $52^{\circ} 13.5{ }^{\prime} \mathrm{N}$ | $13^{\circ} 53,830^{\prime} \mathrm{W}$ |
| 6 | $51^{\circ} 22^{\prime} \mathrm{N}$ | $14^{\circ} 24^{\prime} \mathrm{W}$ |
| 7 | $51^{\circ} 22^{\prime} \mathrm{N}$ | $14^{\circ} 03^{\prime} \mathrm{W}$ |
| 8 | $52^{\circ} 10^{\prime} \mathrm{N}$ | $13^{\circ} 25^{\prime} \mathrm{W}$ |
| 9 | $52^{\circ} 32^{\prime} \mathrm{N}$ | $13^{\circ} 07,500^{\prime} \mathrm{W}$ |
| 10 | $52^{\circ} 43^{\prime} \mathrm{N}$ | $12^{\circ} 55^{\prime} \mathrm{W}$ |
| 11 | $52^{\circ} 43^{\prime} \mathrm{N}$ | $12^{\circ} 43^{\prime} \mathrm{W}$ |
| 12 | $52^{\circ} 38,800^{\prime} \mathrm{N}$ | $12^{\circ} 37^{\prime} \mathrm{W}$ |
| 13 | $52^{\circ} 27^{\prime} \mathrm{N}$ | $12^{\circ} 23^{\prime} \mathrm{W}$ |
| 14 | $52^{\circ} 27^{\prime} \mathrm{N}$ | $12^{\circ} 19^{\prime} \mathrm{W}$ |

By way of derogation from the first subparagraph, transit through the Porcupine Bank while carrying on board the species referred to in that paragraph, shall be permitted in accordance with Article 50(3), (4) and (5) of Regulation (EC) No $1224 / 2009$.

The following TCMs are in place for Nephrops in 7 (excluding 7.a) after EC 850/98 in operation since 2000 :

Minimum Landing Sizes (MLS); total length $>85 \mathrm{~mm}$, carapace length $>25 \mathrm{~mm}$, tail length $>46 \mathrm{~mm}$. Although it is legal to land smaller prawns from this fishery, marketing restrictions imposed by producer organizations in France mean smaller Nephrops ( $<35 \mathrm{~mm}$ CL or 115 mm whole length) are not retained in this fishery.

The mesh size restrictions apply to towed gears in $7 . \mathrm{b}-\mathrm{k}$ targeting Nephrops and are given in Section 7.1. Vessels mainly used 80-99 mm mesh to target Nephrops on the Porcupine Bank.

The landing obligation applied since 2016 for certain vessels that matched the criteria set out in the discard plans:https://ec.europa.eu/fisheries/cfp/fishing_rules/discards_en

### 17.3 Fishery in 2018 and 2019

WGCSE reviewed effort trends for Irish vessels that accounted for over $80 \%$ of the total landings in recent years. The fishery in 2018 took place up to July, after which the fishery was closed, but was reopened in November. The industry reported very good catches of Nephrops but commented that the mean size had declined significantly in 2017 before increased again in 2018.

## Effect of regulations

Prior to 2011 TACs and quotas were applied to the whole of 7 so the FU16 fishery was not been restricted by quotas. Since 2011, the "of which clause" was implemented in the TAC regulation specifically for the Porcupine Bank. Quotas have been very restrictive for Irish vessels and this has led to various changes in fishing patterns. Vessels have tried to optimise the economic value of the catch by targeting areas and periods with relatively smaller ${ }^{1}$ volumes of larger higher value Nephrops. TheFU16 specific quota has also increased the risk of area misreporting, discarding and of highgrading landings. Area misreporting diminished in 2018 with the introduction of a national legislation preventing Irish vessels' fishing in FU16 and other areas during the same fishing trip.

## Information from stakeholders

The provision of grade information by individual fishers and coops remains a highly important assessment input. In 2017 and 2018, the percentage of landings where grade data were provided decreased, but in 2019, it has increased to the highest value in recent years.

| Year | \% of Irish landings where grade data were provided |
| :--- | :--- |
| 2011 | $60 \%$ |
| 2012 | $45 \%$ |
| 2013 | $57 \%$ |
| 2014 | $33 \%$ |
| 2015 | $44 \%$ |
| 2016 | $39 \%$ |
| 2017 | $31 \%$ |
| 2018 | $65 \%$ |
| 2019 |  |

[^10]The industry has also collaborated with the development of the IFSRP survey since 2010 (Stokes and Lordan, 2011).

The Irish industry considers that the stock has increased significantly and no longer requires the Functional Unit "of which" clause.

### 17.4 Data

## InterCatch

Data were available in InterCatch and used on a trial basis.

## Landings

Total international landings increased by $\sim 5 \%$ in 2018 to 2751 t (Figure 20.1 and Table 20.2). Since 2011, total landings for FU16 had included "unallocated landings" from other FU due to misreporting. In 2018, no reallocation was applied as there was no information concerning misreporting.

## Sampling levels

Sampling levels, data aggregating and raising procedures were reviewed by WKNEPH 2013, and are documented in the stock annex. Recent sampling rate is provided in Table 20.3.

Since 2010, landings length distributions have been reconstructed using the methods outlined in the stock annex. This involves using samples of the grade length structure from Irish sampling and estimates of the volume of each commercial size grade provided by the fishing industry. This was used to reconstruct Irish LFDs, landings by other fleets which accounted for $20 \%$ of the landings were unsampled.

## Commercial length-frequency distributions

The time-series of raised international length-frequency distributions of the sampled landings by sex, are given in Figure 20.2. This also shows significant shift towards larger individuals in the landings between 2002-2009 when few individuals at smaller sizes were observed. The length distribution in 2018 is relatively broad and to the right of 2017. The mean lengths by sex and year are presented in Table 20.4.

## Sex ratio

Previous Nephrops working groups have highlighted stability in sex ratio as an important indicator for Nephrops stocks. The landings and fishery-independent survey catches showa dramatic switch in the sex ratio for this stock with larger proportions of females in the catches of 2008 and 2009 (Figure 20.3). Both the commercial and survey data indicate that sex ratio switched back to a more usual situation since 2010 with males accounting for larger proportions of the catch/landings.

Nephrops moult once a year shortly after hatching of eggs in April or May. There is a 24 hour period after moulting when the male Nephrops can mate with the female (Farmer, 1974). If there are insufficient males in the population to mate with the recently moulted females, this can result in a change in female behaviour, whereby unmated females concentrate on feeding and growth instead of reproduction. This so called "sperm limitation" hypothesis could explain the sex ratio
changes observed in the Porcupine Nephrops. WKNEPH 2013 examined the available scientific data on proportions of females mated observed on the Spanish survey. These results showed high proportions of unmated females and a high $L_{50}$ for mated females in catches in 2009. Simulations were also carried out to investigate the densities at which sperm limitation may become an issue given plausible ranges of stock density, sex ratios, search radii. The conclusion was that at the densities recently observed on the Porcupine Bank that sperm limitation was a real possibility.

## Mean weight explorations

The mean weights in in the landings are shown for the full time-series in Figure 20.4 and Table 20.5. In 2019 , the mean weight on the catch samples has been consistent with the grade information in the landings provided by the fishery.

## Discards

There are few historical estimates of discards for this stock. Irish sampling up to 2016, observed very minimal discarding (mainly limited to small and damaged individuals $<5 \%$ by number). Four Irish trips were sampled in 2016. Discards were not recorded on one of these trips. However, on the other three trips, discards were estimated to be around $8 \%, 9 \%$ and $15 \%$ by number $(3 \%, 3 \%$ and $6 \%$ by weight). In 2017, there were two trips where discards were recorded, $17 \%$ and $43 \%$ by number. In 2018, discards were observed on one of the two trips ( $74 \%$ by number), no discards were observed on the other trip. The discarding observed on these trips is likely not reflective of the overall discard pattern as the skippers advised the scientist on board that they had increased their discards to remain within quota during the observed trip. This means that the 2018 discard pattern is unknown, but can be no longer considered negligible. In 2019, discards were observed on one of the three trips ( $14 \%$ by number), and no discards were observed on the other two trips.

A detailed examination of discard estimates was provided in Spain in 2014. No estimate of was provided in InterCatch by Spain since 2015.

## Abundance indices from UWTV surveys

Operational details of the 2019 UWTV survey are available (Aristegui et al., 2019). These surveys use the standard UWTV methodology and conforms to WGNEPS best practice and guidelines. WKNEPH 2013 recommended that these surveys could be used for assessment and provision of catch options. The results are given in Table 20.6. Further detail of the survey is provided in the annex and annual survey reports are available at http://oar.marine.ie/handle/10793/59.

The spatial distributions of burrow densities are shown as a heat maps and bubble plots in Figure 20.5. The 2019 burrow surface shows an area of higher density in the north of the ground. The abundance estimate derived from the krigged burrow surfaces (and adjusted for edge effect) decreased by $9.5 \%$ (from 1117 million in 2018 to 1010 million in 2019), with an estimated area of the ground of $7131 \mathrm{~km}^{2}$ and a CV on the abundance of $5.1 \%$ (Figure 20.6 and Table 20.7).

## Trawl surveys

The longest time-series of fishery-independent source of data is from the Spanish Porcupine trawl survey 2001-2018 (SpPGFS-WIBTS-Q4). This survey is carried out in September when Nephrops catchability is quite low, particularly of adults. Further information on this survey is provided in the IBTS report (ICES, 2015) and in previous IBTS reports.

Distribution of Nephrops catches and biomass in Porcupine surveys between 2001 and 2018 is shown in Figure 20.7. There was a year effect in 2008 when unusual gear parameters were observed. Catch rates in 2011 may also have been reduced due to exceptionally poor weather and gear performance issues. The stratified abundance estimate and biomass increased significantly since 2015 (Figure 20.8).

The size structure of the catches in the survey shows two things: a lower mean size than in the commercial fleets and an increasing trend in mean size for both sexes up to 2008. In 2009, there is large reduction of mean size in both sexes due to a recruiting year class with a modal length at around 27 mm (possibly the 2006 year class). Tracking of cohorts was carried out at WKNEPH 2013 but the results are inconclusive (ICES, 2013). The survey shows increased recruitment since 2013 with significantly increased catch rates of individuals $<21 \mathrm{~mm}$ (Figure 20.9). This has also led to increase catch rates of juveniles and adult Nephrops since 2016.

An Irish Fisheries Science Research Partnership (IFSRP) survey was developed in collaboration with the Irish fishing industry to obtain data from the closed area in 2010-2012. Details of the design and methodology are presented in Stokes and Lordan (2011). The survey uses both commercial gear (Comm) and a baca trawl similar to the SpPGFS-WIBTS-Q4. WKNEPH concluded that the IFSRP trawl survey is too short (with changes in coverage, gears and vessels) to draw an inference about CPUE changes reflecting changing stock abundance (ICES, 2013). The surveys carried out between 2010-2012 provided very useful data on population structure across the ground as well as data on grade structure and maturity-at-length.

## Commercial CPUE

In the past, the Nephrops fishery on the Porcupine Bank was both seasonal and opportunistic with increased targeting during periods of high Nephrops emergence and good weather. Freezing of catches at sea has become increasingly prevalent since 2006 and the fishery now operates throughout the year, mainly targeting larger more valuable Nephrops in lower volumes. Fishing effort has fluctuated considerably in the recent past in response to availability of Nephrops.

Effort and LPUE/CPUE data are generally not standardized, and hence do not take into account vessel capacity, efficiency, seasonality or other factors that may bias perception of LPUE/CPUE and abundance trends over the longer term. WKNEPH concluded that effort and LPUE series should be maintain in the WGCSE report for information purposes (ICES, 2013). WGCSE 2016 recommended presenting the effort in KWDays and LPUE in tonnes/KWDays. Any inferences about changes in stock abundance from these data, should take account of the quality and bias concerns raised above.

These data are presented by country in Table 20.8.

### 17.5 Stock assessment

## Comparison with previous assessments

This assessment is based on UWTV approach outlined in WKNEPH 2013 and using parameter in the stock annex (ICES, 2013). This year's assessment has been updated based on the results of the June 2019 UWTV survey.

## State of the stock

The UWTV results are shown in Table 20.6. In 2017, the harvest rate was above Fmsy for the first time, but in 2018, the harvest rate decreased below FMSY again due to an increase in the abundance
estimate of the 2018 UWTV survey. Total abundance decreased in 2019, but it is still the second highest value in the time-series.

## Catch options table

The inputs to the catch options are given below. At this point, it is not possible to estimate the numbers and mean weights of discards in the fishery, although there are indications that discards have increased since 2016.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Stock abundance (2020) | 1010 million | UWTV survey 2019 (number of individuals). |
| Mean weight in wanted catch | 42.1 g | Average 2017-2019. |
| Mean weight in unwanted catch | - | Unknown. |
| Unwanted catch | - | Discarding is assumed to be negligible. |
| Discards survival | - | Not applicable. |
| Dead unwanted catch | Assumed to be zero. |  |

### 17.6 Reference points

New reference points were evaluated by WKMSYREF4 (ICES, 2016a) and advised by ICES (2016b). The Fmsy for this stock was increased from $5.0 \%$ to $6.2 \%$. The Fmsy for this stock is based on $\mathrm{F}_{0.1}$ for both sexes combined given the low density of Nephrops on the Porcupine Bank.

| Stock code | MSY Flower* | F $_{\text {MSY }}{ }^{*}$ | MSY Fupper $^{*}$ with AR | MSY B trigger | MSY F upper $^{*}$ with no AR |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| nep-16 | $5.0 \%$ | $6.2 \%$ | $6.2 \%$ | Not defined | $6.2 \%$ |

* Harvest rate (HR).


### 17.7 Management strategies

There is no management plan for this stock.
The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters(EU, 2019). This plan applies to Norway lobster (Nephrops norvegicus) by functional unit in ICES Subarea 7 and also demersal stocks.

### 17.8 Quality of assessment and forecast

The main quality considerations for this stock are related to mean weight and discarding. The mean weight for this stock has been fluctuating, the most recent estimates maybe overestimate due to the non-inclusion of discards. The mean weight has declined in the last fewyears as strong year classes recruit to the fishery. Since 2017, a recent mean weight in the landing was considered the most appropriate basis in the calculation of catch scenarios. In previous years, a long-term mean weight was used.

The provision of grade information (commercial size categories) by individual fishers and fishery coops remains highly important for calculating mean weight in the landings. The proportion
of landings for which grade data were provided has declined from $49 \%$ in 2016 to $20 \%$ in 2018, increasing the uncertainty around the mean weight assumption.

There is good evidence from surveys and length structure of landings that recruitment has improved and this has resulted in a reduction in mean weight in the stockin recent years. The mean weight in 2019 may well increase again as the stronger cohorts grows. Currently there is no methodology to take this into account in the calculation of catch options.

Up to 2015, discarding was considered negligible for this functional unit. Since 2016, the amount of discards observed on catch sampling trips has increased. This may be temporary linked to the incoming recruitment. It will result is an underestimate of recent harvest rates of similar magnitude to the numbers. The main concern is that the mean weight derived from the landings grades maybe bias due to unknown discarding levels.
The UWTV survey provides abundance since 2012 (except 2015) with high precision, but the time-series is short and an abundance MSYB ${ }_{\text {trigger }}$ has yet to be defined.
The landings are considered fairly well estimated up to 2018 (an unallocated component related to area misreporting was included from 2011 to 2017).

### 17.9 Recommendation for next benchmark

This stock was benchmark in 2013 at WKNEPH. WGCSE will keep the stock under close review and recommend future benchmark as required.

### 17.10 Management considerations

There is a separate catch limit for Functional Unit (FU) 16 within the wider TAC for Subarea 7. This has resulted in very restrictive quotas for some vessels, which increased area misreporting and the risk of discarding from 2011 to 2017. Area misreporting diminished in 2018 with the introduction of a national legislation restricting Irish vessels' fishing areas, where since March 2018, Irish vessels targeting Nephrops in subareas 6 and 7 may only fish in either of (1) Subarea 6 or Subarea 7, excluding FU16, or (2) FU16 of Subarea 7 (Fisheries Management Notice No. 20 of 2018). Although there was an increase on landings in 2018, the harvest rate decreased due to an increase in the abundance estimate of the 2018 UWTV survey. Given the vulnerability of this stock to overexploitation, the separate catch limit for Functional Unit (FU) 16 should remain in place.

A seasonal closed area (May 1-July 31) has been in place since 2010. The period of the closure was reduced to one month, May, since 2013. There hasn't been an ICES evaluation of the impact of this closure and whether it provides a conservation benefit over and above catch limits. Some sectors of the fishing industry want to extend the period of closure because they believe that this is a more effective conservation measure than catch limits.

Productivity of deep-water Nephrops stocks is generally lower than that in shelf waters, though individual Nephrops grow to relatively large sizes and attain high market prices. Other deepwater Nephropsstocks off the Spanish and Portuguese coast have collapsed and have been subject to recovery measures for several years e.g. FU25, 26, 27 and 31. Recruitment in Nephrops populations in deep water may be more sporadic than for shelf stocks with strong larval retention mechanisms. This makes these stocks more vulnerable to over exploitation and potential recruitment failure as has been observed on the Porcupine Bank in the early 2000s.

From 2019, vessels using highly selective gears in Subarea 7 can be exempted from the landings obligation on the basis of the high survival exemption (see discard plans). It is unknown if Nephrops discarded on the Porcupine Bank could actually survive the discarding process.

Discarding by the Nephrops trawl fishery is around $50 \%$ of the total catch by weight. The main species that are discarded by weight are blue mouth-red fish, blue whiting and argentines (Marine Institute and Bord Iascaigh Mhara, 2011).

### 17.11 References

Aristegui, M., O’Brien, S., Tully, D., Galligan, S., McCorriston, P., Bentley, K. and Lordan, C. 2019. Porcupine Bank Nephrops Grounds (FU16) 2019 UWTV Survey Report and catch scenarios for 2020. Marine Institute UWTV Survey report.

EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. Official Journal of the European Union, L 83: 1-17. http://data.europa.eu/eli/reg/2019/472/oj.
ICES.2013. Report of the Benchmark Workshop on Nephrops Stocks(WKNEPH), 25 February-1 March 2013, Lysekil, Sweden. ICES CM 2013/ACOM:45. 230 pp.
ICES. 2015. Report of the International Bottom Trawl Survey Working Group (IBTSWG), 23-27 March 2015, Bergen, Norway. ICES CM 2015/SSGIEOM:24. 278 pp.

ICES. 2016a. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMS YREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.
ICES. 2016b. EU request to ICES to provide Fmsy ranges for selected stocks in ICES subareas 5 to 10. In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.2.3.1.

Marine Institute and Bord Iascaigh Mhara. 2011. Atlas of Demersal Discarding, Scientific Observations and Potential Solutions, Marine Institute, Bord Iascaigh Mhara, September 2011. ISBN 978-1-902895-50-5. 82 pp. http://hdl.handle.net/10793/666.
Stokes, D. and Lordan, C. 2011. "Irish fisheries-science research partnership trawl survey of the Porcupine Bank NephropsGrounds July 2010", Irish Fisheries Bulletin No. 39, Marine Institute 2011. http://hdl.handle.net/10793/712.

Table 20.1. Nephrops Porcupine Bank (FU 16): Of which catch limit.

| Year | France | Ireland | Spain | Total |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 241 | 454 | 377 | 188 | 1260 |
| 2012 | 238 | 457 | 380 | 185 | 1260 |
| 2013 | 340 | 653 | 543 | 264 | 1800 |
| 2014 | 349 | 671 | 557 | 271 | 1848 |
| 2015 | 349 | 671 | 558 | 272 | 1850 |
| 2016 | 586 | 1124 | 935 | 455 | 3100 |
| 2018 | 516 | 959 | 825 | 401 | 273 |
| 2019 | 500 | 798 |  | 268 |  |

Table 20.2. Nephrops Porcupine Bank (FU 16): Landings (tonnes) by country.

| Year | France | Ireland | Spain | UK E\& W | UK Scotland | Unallocated | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 514 |  |  |  |  |  | 514 |
| 1966 | 0 |  |  |  |  |  | 0 |
| 1967 | 441 |  |  |  |  |  | 441 |
| 1968 | 441 |  |  |  |  |  | 441 |
| 1969 | 609 |  |  |  |  |  | 609 |
| 1970 | 256 |  |  |  |  |  | 256 |
| 1971 | 500 |  | 1444 |  |  |  | 1944 |
| 1972 | 0 |  | 1738 |  |  |  | 1738 |
| 1973 | 811 |  | 2135 |  |  |  | 2946 |
| 1974 | 900 |  | 1894 |  |  |  | 2794 |
| 1975 | 0 |  | 2150 |  |  |  | 2150 |
| 1976 | 6 |  | 1321 |  |  |  | 1327 |
| 1977 | 0 |  | 1545 |  |  |  | 1545 |
| 1978 | 2 |  | 1742 |  |  |  | 1744 |
| 1979 | 14 |  | 2255 |  |  |  | 2269 |
| 1980 | 21 |  | 2904 |  |  |  | 2925 |
| 1981 | 66 |  | 3315 |  |  |  | 3381 |
| 1982 | 358 |  | 3931 |  |  |  | 4289 |
| 1983 | 615 |  | 2811 |  |  |  | 3426 |
| 1984 | 1067 |  | 2504 |  |  |  | 3571 |
| 1985 | 1181 |  | 2738 |  |  |  | 3919 |
| 1986 | 1060 |  | 1462 | 69 |  |  | 2591 |
| 1987 | 609 |  | 1677 | 213 |  |  | 2499 |
| 1988 | 600 |  | 1555 | 220 |  |  | 2375 |
| 1989 | 324 | 350 | 1417 | 24 |  |  | 2115 |
| 1990 | 336 | 169 | 1349 | 41 |  |  | 1895 |
| 1991 | 348 | 170 | 1021 | 101 |  |  | 1640 |
| 1992 | 665 | 311 | 822 | 217 |  |  | 2015 |


| Year | France | Ireland | Spain | UK E\& W | UK Scotland | Unallocated | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 799 | 206 | 752 | 100 |  |  | 1857 |
| 1994 | 1088 | 512 | 809 | 103 |  |  | 2512 |
| 1995 | 1234 | 971 | 579 | 152 |  |  | 2936 |
| 1996 | 1069 | 508 | 471 | 182 |  |  | 2230 |
| 1997 | 1028 | 653 | 473 | 255 |  |  | 2409 |
| 1998 | 879 | 598 | 405 | 273 |  |  | 2155 |
| 1999 | 1047 | 609 | 448 | 185 |  |  | 2290 |
| 2000 | 351 | 227 | 213 | 120 |  |  | 910 |
| 2001 | 425 | 369 | 270 | 158 |  |  | 1222 |
| 2002 | 369 | 543 | 276 | 139 |  |  | 1327 |
| 2003 | 131 | 307 | 489 | 108 | 29 |  | 1064 |
| 2004 | 289 | 494 | 468 | 126 | 28 |  | 1406 |
| 2005 | 397 | 754 | 681 | 208 | 156 |  | 2197 |
| 2006 | 462 | 731 | 636 | 201 | 155 |  | 2185 |
| 2007 | 302 | 1060 | 384 | 146 | 183 |  | 2074 |
| 2008 | 26 | 562 | 234 | 41 | 138 |  | 1000 |
| 2009 | 4 | 356 | 348 | 13 | 159 |  | 879 |
| 2010 | 4 | 579 | 240 | 10 | 90 |  | 922 |
| 2011 | 8 | 643 | 182 | 23 | 122 | 301 | 1278 |
| 2012 | 0.46 | 605 | 198 | 0 | 134 | 320 | 1258 |
| 2013 | 5.8 | 651 | 132 | 1 | 118 | 234 | 1141 |
| 2014 | 3 | 813 | 129 | 0 | 96 | 148 | 1189 |
| 2015 | 3 | 744 | 84 | 0 | 109 | 454 | 1394 |
| 2016 | 35 | 1052 | 58 | 1 | 160 | 849 | 2154 |
| 2017 | 63 | 743 | 73 | 249 | 131 | 1373 | 2632 |
| 2018 | 81 | 2079 | 158 | 288 | 144 | 0 | 2751 |

Table 20.3. Nephrops Porcupine Bank (FU 16): Recent sampling used in the assessment.

| Year | Spain |  | France |  | Ireland |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Trips | Type | Number of Trips | Type | Number of Trips | Type |
| 2018 |  |  |  |  | 2 | Graded Landings |
| 2017 |  |  |  |  | 2 | Graded Landings |
| 2016 |  |  |  |  | 4 | Graded Landings |
| 2015 |  |  |  |  | 3 | Graded Landings |
| 2014 |  |  |  |  | 3 | Graded Landings |
| 2013 |  |  |  |  | 3 | Graded Landings |
| 2012 | 0 |  | 0 |  | 3 | Graded Landings |
| 2011 | 0 |  | 0 |  | 2 | Graded Landings |
| 2010 | 0 |  | 0 |  | 3 | Graded Landings |

Table 20.4. Nephrops Porcupine Bank (FU 16): Mean sizes (mm CL) of male and female Nephrops in Spanish, French and Irish landings and the Spanish Porcupine Groundfish survey 1981-2018.

| Year | Spain |  | Ireland |  | France |  | Porcupine Survey |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Landings |  | Landings |  | Landings |  | Catch |  |
|  | Males | Females | Males | Females | Males | Females | Males | Females |
| 1981 | 39.9 | 34.5 | - | - | - | - | - | - |
| 1982 | 40.9 | 34.8 | - | - | - | - | - | - |
| 1983 | 40.8 | 34.0 | - | - | - | - | - | - |
| 1984 | 39.7 | 33.1 | - | - | - | - | - | - |
| 1985 | 38.7 | 33.5 | - | - | - | - | - | - |
| 1986 | 40.7 | 36.4 | - | - | - | - | - | - |
| 1987 | 39.3 | 35.0 | - | - | - | - | - | - |
| 1988 | 40.7 | 38.3 | - | - | - | - | - | - |
| 1989 | 40.5 | 36.8 | - | - | - | - | - | - |
| 1990 | 41.0 | 36.1 | - | - | - | - | - | - |
| 1991 | 39.4 | 34.5 | - | - | - | - | - | - |
| 1992 | 39.2 | 34.1 | - | - | - | - | - | - |
| 1993 | 41.6 | 36.1 | - | - | - | - | - | - |
| 1994 | 40.8 | 36.5 | - | - | - | - | - | - |
| 1995 | 41.3 | 36.6 | 40.7 | 36.5 | 43.2 | 38.3 | - | - |
| 1996 | 41.6 | 35.1 | 34.6 | 35.3 | 41.7 | 38.9 | - | - |
| 1997 | 39.7 | 34.8 | 35.9 | 34.5 | 41.9 | 38.4 | - | - |
| 1998 | 41.1 | 34.6 | 37.2 | 35.6 | 41.9 | 38.4 | - | - |
| 1999 | 41.5 | 35.7 | 36.6 | 33.7 | 43.1 | 39.1 | - | - |
| 2000 | 41.1 | 34.8 | na | na | 45.3 | 40.5 | - | - |
| 2001 | 41.1 | 36.3 | 37.8 | 35.4 | 45.4 | 39.4 | 36.0 | 28.9 |
| 2002 | 39.7 | 35.3 | 36.1 | 38.5 | 45.3 | 40.3 | 37.5 | 31.7 |
| 2003 | 41.4 | 37.8 | 44.5 | 36.2 | 46.2 | 38.9 | 39.7 | 30.9 |
| 2004 | 43.5 | 38.5 | 43.5 | 35.7 | 46.4 | 41.5 | 39.9 | 30.5 |
| 2005 | 43.4 | 38.1 | 46.9 | 40.6 | 45.9 | 41.0 | 45.1 | 33.8 |
| 2006 | 43.9 | 38.0 | na | na | 48.9 | 41.4 | 44.3 | 35.0 |


| Year | Spain |  | Ireland |  | France |  | Porcupine Survey |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Landings |  | Landings |  | Landings |  | Catch |  |
|  | Males | Females | Males | Females | Males | Females | Males | Females |
| 2007 | 43.7 | 41.0 | na | na | 48.3 | 43.8 | 45.9 | 37.8 |
| 2008 | 51.0 | 40.6 | 43.3 | 37.5 | na | na | 48.8 | 38.7 |
| 2009 | 43.0 | 42.7 | 44.1 | 40.1 | na | na | 32.6 | 28.9 |
| 2010 | na | na | 43.2 | 40.4 | na | na | 36.3 | 31.8 |
| 2011 | na | na | 39.5 | 38.4 | na | na | 39.0 | 33.6 |
| 2012 | na | na | 41.1 | 38.1 | na | na | 41.1 | 30.8 |
| 2013 | na | na | 42.9 | 38.9 | na | na | 37.6 | 25.1 |
| 2014 | na | na | 45.1 | 40.9 | na | na | 36.4 | 31.0 |
| 2015 | na | na | 40.3 | 39.7 | na | na | 35.5 | 32.7 |
| 2016 | na | na | 37.8 | 37.3 | na | na | 32.2 | 27.8 |
| 2017 | na | na | 35.7 | 32.9 | na | na | 34.1 | 26.8 |
| 2018 | na | na | 39.7 | 36.2 | na | na | 35.0 | 28.1 |

Table 20.5. Nephrops Porcupine Bank (FU16): Time-series of numbers landed and mean weight in the landings.

| Year | Numbers (millions) | Weight Landed (Tonnes) | Mean Weight in landings (gr) |
| :---: | :---: | :---: | :---: |
| 1986 | 55.7 | 2591 | 46.53 |
| 1987 | 60.3 | 2499 | 41.42 |
| 1988 | 48.1 | 2375 | 49.34 |
| 1989 | 45.6 | 2115 | 46.4 |
| 1990 | 38.9 | 1895 | 48.67 |
| 1991 | 37.3 | 1640 | 43.98 |
| 1992 | 47 | 2015 | 42.84 |
| 1993 | 38.5 | 1857 | 48.29 |
| 1994 | 54.4 | 2512 | 46.15 |
| 1995 | 65.5 | 2936 | 44.79 |
| 1996 | 52.9 | 2230 | 42.15 |
| 1997 | 59.1 | 2409 | 40.73 |
| 1998 | 49.9 | $2155$ | 43.16 |
| 1999 | 52.3 | 2290 | 43.76 |
| 2000 | 15.1 | 910 | 60.13 |
| 2001 | 24.6 | 1222 | 49.65 |
| 2002 | 32 | 1327 | 41.49 |
| 2003 | 18.4 | 1064 | 57.76 |
| 2004 | 21.5 | $1406$ | $65.28$ |
| 2005 | 31.5 | 2197 | 69.84 |
| 2006 | 28.7 | 2185 | 76.24 |
| 2007 | 29.2 | $2074$ | 71.05 |
| 2008 | 17.9 | 1000 | 55.89 |
| 2009 | 16.5 | 879 | 53.19 |
| 2010 | 14.1 | $922$ | 65.32 |
| 2011 | 27.9 | 1278 | 45.81 |
| 2012 | 25 | 1258 | 50.36 |
| 2013 | 19.8 | 1141 | 57.54 |
| 2014 | 17.3 | 1189 | 68.54 |
| 2015 | 27.4 | 1394 | 50.86 |
| 2016 | 53.5 | 2154 | 40.29 |
| 2017 | 84.9 | 2632 | 31.01 |
| 2018 | 66.2 | 2751 | 41.55 |
| 2019 |  |  | 53.73 |
| Average 2017-2019 |  |  | 42.10 |

Table 20.6. Nephrops Porcupine Bank (FU16): Assessment summary.

|  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Millions |  |  |  |  | \% | tonnes |  | \% |  | grammes |  |
| 2012 | 787 | 79 | 25 | 0 | 25 | 3.2 | 1258 | 0 | 0 | 0 | 50.4 | NA |
| 2013 | 768 | 61 | 20 | 0 | 20 | 2.6 | 1141 | 0 | 0 | 0 | 57.5 | NA |
| 2014 | 722 | 35 | 17 | 0 | 17 | 2.4 | 1189 | 0 | 0 | 0 | 68.5 | NA |
| 2015 | NA | NA | 27 | 0 | 27 | 3.3** | 1394 | 0 | 0 | 0 | 50.9 | NA |
| 2016 | 958 | 68 | 53 | NA | 53 | 5.6 | 2154 | NA | NA | NA | 40.3 | NA |
| 2017 | 850 | 90 | 85 | NA | 85 | 10.0 | 2632 | NA | NA | NA | 31.0 | NA |
| 2018 | 1117 | 92 | 66 | NA | 66 | 5.9 | 2751 | NA | NA | NA | 41.6 | NA |
| 2019 | 1010 | 101 |  |  |  |  |  |  |  |  | 53.7 |  |

*Discarding up to 2015 was considered to be negligible. Discard estimates are not available since 2016 and are therefore not included in the assessment.
** The harvest rate is estimated based on a linear interpolation of abundance for 2015 as no survey was carried out in this year.
*** Values since 2016 onwards may be underestimates due to insufficient discard data.
NA = not available.

Table 20.7. Nephrops Porcupine Bank (FU16): Results summary table for geostatistical analysis of UWTV survey series.

| YearNumber of <br> stations | Mean Density ad- <br> justed (burrow/m $)$ | Domain Area <br> $\left(\mathrm{km}^{2}\right)$ | Geostatistical Abundance Esti- <br> mate adjusted (millions) | CV on Burrow <br> estimate (\%) |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 2012 | 47 | 0.151 | 7108 | 787 | 4.9 |
| 2013 | 68 | 0.106 | 7108 | 768 | 4.4 |
| 2014 | 67 | 0.099 | 7108 | 722 | 2.5 |
| 2015 | 0 | 0.132 | 7108 | 858 | 3.6 |
| 2016 | 65 | 0.118 | 7134 | 1117 | 5.4 |
| 2017 | 63 | 0.156 | 7130 | 1010 | 4.2 |
| 2018 | 69 |  |  |  |  |

Table 20.8. Nephrops Porcupine Bank (FU16): Effort and LPUE for the various different fleets exploiting the stock 1971-2017.

| Spain ${ }^{1}$ |  | France ${ }^{2}$ |  | Ireland ${ }^{3}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { Effort ('000's } \\ & \text { Hrs) } \end{aligned}$ | LPUE <br> (kg/hr) | $\begin{aligned} & \text { Effort² ('000's } \\ & \text { Hrs) } \end{aligned}$ | LPUE (>10\%) (kg/hr) |  | Effort ${ }^{3}$ ('000's KwDays) | LPUE <br> (t/KWdays) |
| 1980 | 318 | 9 |  |  |  |  |
| 1981 | 272 | 12 |  |  |  |  |
| 1982 | 237 | 17 |  |  |  |  |
| 1983 | 196 | 14 | 18 | 35 |  |  |
| 1984 | 194 | 13 | 30 | 35 |  |  |
| 1985 | 200 | 14 | 33 | 36 |  |  |
| 1986 | 162 | 9 | 28 | 38 |  |  |
| 1987 | 174 | 10 | 24 | 26 |  |  |
| 1988 | 180 | 9 | 22 | 27 |  |  |
| 1989 | 173 | 8 | 14 | 23 |  |  |
| 1990 | 159 | 9 | 15 | 23 |  |  |
| 1991 | 138 | 7 | 19 | 18 |  |  |
| 1992 | 96 | 9 | 32 | 21 |  |  |
| 1993 | 80 | 9 | 36 | 22 |  |  |
| 1994 | 80 | 10 | 38 | 28 |  |  |
| 1995 | 67 | 9 | 42 | 30 | 584.9 | 1.4 |
| 1996 | 58 | 8 | 41 | 26 | 192.5 | 1.59 |
| 1997 | 57 | 8 | 41 | 25 | 327.3 | 1.26 |
| 1998 | 56 | 7 | 40 | 22 | 284.6 | 1.59 |
| 1999 | 53 | 8 | 43 | 21 | 278 | 1.29 |
| 2000 | 47 | 5 | 23 | 14 | 92.8 | 1.25 |
| 2001 | 44 | 6 | 24 | 15 | 230.2 | 1.12 |
| 2002 | 54 | 5 | 18 | 18 | 339.8 | 1.3 |
| 2003 | 66 | 5 | 7 | 19 | 294.7 | 0.8 |
| 2004 | 59 | 10 | 9 | 25 | 569.2 | 0.68 |
| 2005 | 60 | 13 | 15 | 26 | 756.2 | 0.83 |


| Spain ${ }^{1}$ |  | France ${ }^{2}$ |  |  | Ireland ${ }^{3}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { Effort ('000's } \\ & \text { Hrs) } \end{aligned}$ | LPUE <br> (kg/hr) | $\begin{aligned} & \text { Effort² ('000's } \\ & \text { Hrs) } \end{aligned}$ | $\begin{aligned} & \text { LPUE (>10\%) } \\ & \text { (kg/hr) } \end{aligned}$ |  | Effort ${ }^{3}$ ('000's KwDays) | LPUE <br> (t/KWdays) |
| 2006 | 65 | 9 | 22 | 21 | 952.8 | 0.72 |
| 2007 | 58 | 8 | 17 | 18 | 1199.4 | 0.81 |
| 2008 | 42 | 6 | 4 | 7 | 830.7 | 0.67 |
| 2009 | 44 | 7 |  |  | 411.3 | 0.83 |
| 2010 | 42 | 6 |  |  | 704.1 | 0.81 |
| 2011 | na | na |  |  | 986.9 | 0.63 |
| 2012 | 15 | na |  |  | 817.1 | 0.63 |
| 2013 | na | na |  |  | 885.7 | 0.92 |
| 2014 | na | na |  |  | 1019.8 | 0.92 |
| 2015 | na | na |  |  | 1219.2 | 0.99 |
| 2016 | na | na |  |  | 1359.3 | 1.43 |
| 2017 | na | na |  |  | 1328.9 | 1.59 |
| 2018 | na | na |  |  | 1721.2 | 1.21 |

1 = Effort and LPUE between 1980 and 2010 was estimated based on fishing days in 7 . Effort in 2012 was based on logbooks for FU16.
$2=$ Effort and LPUE for vessels where $<\mathbf{1 0 \%}$ of landed value was Nephrops.
3 = Effort and LPUE for vessels where $30 \%$ of the landed weight was Nephrops.


Figure 20.1. Nephrops in FU16 (Porcupine Bank). WG's best estimates of landings in tonnes by country.


Figure 20.2. Nephrops in FU16 (Porcupine Bank). Female and male length distributions of raised international landings. Vertical dashed lines refer to Minimum Landing Size ( $\mathbf{2 5} \mathrm{mm}$ ).


Figure 20.3. Nephrops in FU16 (Porcupine Bank). The percentage males in the landings and survey over time.


Figure 20.4. Nephrops in FU16 (Porcupine Bank). Mean weight in the commercial landings.


Figure 20.5. Nephrops in FU16 (Porcupine Bank). Contour plots of the krigged density estimates for the Porcupine UWTV surveys from 2012 (top left) to 2019 (bottom left) (except 2015).


Figure 20.6. Nephrops in FU16 (Porcupine Bank). Nephrops abundance estimates for 2012-2019 (except 2015).


Figure 20.7. Nephrops in FU16 (Porcupine Bank). Distribution of Nephrops norvegicus in Porcupine surveys top biomass, bottom No. juveniles ( $<\mathbf{2 0} \mathbf{~ m m}$ carapace length).


Figure 20.8. Nephrops in FU16 (Porcupine Bank). Changes in Nephrops norvegicus biomass and number stratified indices during Porcupine Survey time-series. Boxes mark parametric standard error of the stratified abundance index. Lines mark bootstrap confidence intervals ( $\alpha=0.80$, bootstrap iterations=1000).


Figure 20.9. Nephrops in FU16 (Porcupine Bank). Abundance of a) small Nephrops (<21 mm), b) juveniles between $21-26 \mathrm{~mm}$ and c) adults ( $>26 \mathrm{~mm}$ ) in Porcupine survey.


Figure 20.10. Nephrops in FU16 (Porcupine Bank). Summary of stock status for Porcupine Nephrops.

# 18 Norway lobster (Nephrops norvegicus) in Division 7.b, Functional Unit 17 (west of Ireland, Aran Grounds) 

## Type of assessment in 2019

This stock was inter-benchmarked in September 2015 by correspondence (ICES, 2015). The assessment and catch options follow the agreed procedures set out in the stock annex.

## ICES advice applicable to 2018

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2014-2016, catches in 2018 should be no more than 551 tomnes.

To ensure that the stock in functional unit (FU) 17 is exploited sustainably, management should be implemented at the functional unit level."

## ICES advice applicable to 2019

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2015-2017, catches in 2019 should be no more than 1002 tonnes.

To ensure that the stock in Functional Unit 17 is exploited sustainably, management should be implemented at the functional unit level."

### 18.1 General

## Stock description and management units

The Aran Grounds Nephrops stock (FU17) covers ICES rectangles 34-35 D9-E0 within 7.b. This stock is included as part of the TAC Area 7 Nephrops, which includes the following stocks: Irish Sea East and West (FU14, FU15), Porcupine Bank (FU16), northwesternIrish Coast(FU18), southeastern and southwestern Irish Coast (FU19) and the Celtic Sea (FU20-22).

Map below shows FU17 assessment area (blue) and TAC area (red). See Section 18 for details on NephropsSubarea 7 general section.


## Ecosystem aspects

Details of the ecosystem on the Aran grounds are provided in the stock annex updated by IBPNeph (ICES, 2015a).

## Fishery description

A description of the fleet is given in the stock annex. The time-series of numbers of vessels is updated in Figure 21.1.1. The numbers of vessels has been relatively stable since 1995. The timeseries of vessel power is shown as a box and kite plot in Figure 21.1.2.
The majority of the landings are made with 80 mm mesh.
The majority of the landings come from the grounds to the west and southwest of the Aran Islands known as the 'back of the Aran ground' (Seestock annex). The fishery on the Aran Grounds operates throughout the year, weather permitting with a seasonal trend (See stock annex).

## Fishery in 2018

In recent years, several newer vessels specializing in Nephrops fishing have participated periodically in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates. There has been a trend for Irish vessels to switch to multi (quad)rig trawls since 2012. These vessels are more efficient atcatching Nephrops (BIM, 2014).

## Information from stakeholders

Voluntary effort restriction were put in place by the Irish fishing industry in April and May 2015. These measures reduced catches and effort significantly on the stock in advance of the 2015 UWTV survey.

### 18.2 Data

## InterCatch

Data were available in InterCatch and used for catch data only.

## Landings

The reported landings time-series is shown in Figure 21.2.1 and Table 21.2.1. The 2018 landings increased by about $82 \%$ from those made in 2017 and amounted to 536 t .

## Effort

The IBPNeph 2015 reviewed Irish commercial landings and effort data in detail. They concluded that effort should be reported in the WGCSE report in KWdays and LPUE should be reported in KG/kwdays in the knowledge that the trend is likely to be a biased underestimate because it is not adjusted for efficiency or behavioural changes. The time-series of effort and LPUE is updated in Figure 21.2.2 and Table 21.2.2. There was a significant decline in LPUE and effort in 2015 due to the local management efforts put in place in April and May. In 2016 effort level increased to values similar to those observed previously prior to 2011.However, in 2017 and 2018 effort levels declined again to values similar to 2015.

## Sampling levels

Sampling levels, data aggregating and raising procedures were reviewed by IBPNeph 2015 and are documented in the stock annex. The time-series of samples is shown in Figure 21.2.3 and Table 21.2.3. Sampling levels in 2018 were similar to 2016 and 2017 levels.

## Commercial length-frequency distributions

The raised catch length distributions are shown in Figure 21.2.4. The mean length of females decreased in 2018, increasing the discard rate for females.

## Sex ratio

In 2018, due to the strong dominance of small females in the catches (Figure 21.2.4), the difference on the proportion of males between the catches and the landings is higher than in previous years (Figure 21.2.5). Sex ratio has a distinct seasonal pattern with lowest male proportions in the samples in May and June. Males dominate the catches in the autumn and winter.

## Mean weight explorations

Explorations of the mean weight in the catch samples by sex shows a strong cyclical pattern in females, which corresponds with the emergence of mature females from the burrows to mate in summer (Figure 21.2.6). The annual mean weight estimate for landings and discards is shown in Figure 21.2.7. The mean weight estimates have been relatively stable from 2011, where main change occurred in 2008-2011. There has been an increase on the landings mean weight since 2015.

## Discarding

Table 21.2.4 gives weights, numbers and proportions of the landings and discard raised internationally according to the stock annex. A $25 \%$ discard survival rate is assumed in line with other Nephrops stocks in the Celtic sea (see stock annex) as the basis for the catch scenarios. Gear selectivity trials by Bord Iascaigh Mhara (BIM, 2017) reported a $64 \%$ survivor rate for Nephrops caught in a trawl with a SELTRA selectivity device in the outer Galway Bay area.

## Abundance indices from UWTV surveys

The spatial extent of the Nephropsgrounds in FU17 was re-defined by IBPNeph 2015 and the total abundance estimates were revised using a new procedure (ICES, 2015a). The redefinition of the polygons in FU17 resulted in $\sim 30 \%$ increase in overall area from $1007 \mathrm{~km}^{2}$ to $1320 \mathrm{~km}^{2}$ (stock annex). Operational details of the 2019 UWTV survey are available (Aristegui et al., 2019).

The spatial distributions of burrow densities are shown in Figure 21.2.8. The densities have fluctuated considerably over the time-series and throughout the Aran grounds. In general, the densities are higher towards the middle-western side of the ground and there is a notable trend towards lower densities towards the east. On the southwestern boundary, there are often high densities close to the boundary. In this area, there is a sharp transition from mud to rocky substrate. The decrease in densities in 2019 was mainly towards the north of the ground.

In 2018, the Aran Grounds account for $\sim 93 \%$ of the total estimated burrow abundance from FU17 (Table 21.2.5). Galway Bay accounts for $\sim 5 \%$ and Slyne Head for $\sim 2 \%$ (Table 21.2.6). The Galway Bay estimates fluctuate widely but appear to be highly correlated with the Aran ground except in 2004 (Figure 21.2.9). Estimates for theSlyneHead ground alsofluctuate considerably but show no significant correlation with the other areas except for the peaks of 2010, 2015 and 2018 (Figure 21.2.9). In 2019, abundance estimates decreased for the three grounds.

Aran ground abundance estimate's CV (4\%) (Table 21.2.5) is well below the recommendation of $20 \%$ by SGNEPS (ICES, 2012). The CVs on the abundance estimates for Galway Bay and Slyne Head are relatively higher ( $11 \%$ and $8 \%$ ) (Table 21.2.6), but still within the recommendation, showing the surveys are precise. Figure 21.2.10 and Table 21.2 .7 show the total abundance estimate for FU17 with the IBPNeph proposed MSY B ${ }_{\text {trigger. }}$. The 2019 combined abundance estimate ( 493 million) was $11 \%$ lower than in 2018 and is below the MSY $B_{\text {trigger }}$ ( 540 million).

### 18.3 Assessment

## Comparison with previous assessments

The WGCSE 2019 carried out an UWTV based assessment for this stock. The methods used were very much in line with WKNEPH (ICES, 2009) and the approach taken for other Nephrops stocks in 6 and 7 by WGCSE. This approach was inter-benchmarked at IBPNeph (ICES, 2015a).

## State of the stock

UWTV abundance estimates suggest that the stock size has fluctuated widely with an overall declining trend and is below MSY Btrigger since 2012 (except 2015 and 2018). The 2019 estimate is the sixth lowest observed in the time-series and is below the MSY Btrigger. The 2019 abundance remains below the average of the series (geomean [2002-2019]: 655 million). Harvest rate (calculated as (landings + dead discards)/abundance estimate) has been below the FmsYproxy for the last two years (Table 21.3.1.and Figure 21.3.1).

### 18.4 Catch scenario table

Catch scenario table inputs and historical estimates of mean weight in landings and harvest ratios are presented in Table 21.3.1, and summarised below. The calculation of catch options for the Aran Grounds follows the procedure outlined in the stock annex.

The basis for the catch scenarios.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Stock abundance | 493 million indi- <br> viduals | UWTV Survey 2019 |
| Mean weight in <br> landings | 22.4 g | Average 2008-2018 |
| Mean weight in <br> discards | 11.04 g | Average 2008-2018 |
| Discard rate | $23.6 \%$ | Average (proportion by number) 2016-2018. Calculated as discards/(landings + <br> discards) |
| Discard survival <br> rate | Only applies in scenarios where discarding is allowed. |  |
| Dead discard rate | Average 2016-2018 (proportion by number). Calculated as dead discards di- <br> vided by dead removals (landings + dead discards). Only applies in scenarios <br> where discarding is allowed. |  |

Given the fluctuations observed in mean weights for landings and discards (Figure 21.2.7) an average from 2008 to the most recent year is used in the calculation of catch options as set out in the stock annex. The discard rates and proportions for the last three years are used to account for recent on-board retention practices (this is also according to the stock annex).

### 18.5 Reference points

New reference points were defined for this stock at the IBPNeph (ICES, 2015a) and no new proposals were made by WKMSYRef4 (ICES, 2016a). For Nephrops stocks MSY Btrigger has been defined as the lowest stock size from which the abundance has increased. This corresponds to the abundance observed in 2008 rounded to the nearest $10=540$ million individuals (Figure 21.2.10 and Table 21.2.7).

The Fmsy proxy was revised during the benchmark in 2015. The observed burrow density has declined, from high ( $>0.8$ individuals $\mathrm{m}^{-2}$ ) at the start of the series to medium density ( $\sim 0.3$ individuals $\mathrm{m}^{-2}$ ) towards the end of the time-series. The nature of the fishery has also changed, from a continuous fishery throughout the year to a fishery, which is more concentrated on sporadic periods of high catch rates. For these reasons a harvest rate consistent with a combined sex $\mathrm{F}_{0.1}=$ $8.5 \%$ is considered an appropriate proxy for $\mathrm{F}_{\text {MSY }}$.

These should remain under review by WGCSE and may be revised while data become available.

### 18.6 Management strategies

As yet there are no explicit management strategies for this stock, but there have been some discussions among the fishing industry and scientists about developing a long-term plan for the
management of the Aran fishery. Sustainable utilization of the Nephrops stock will form the cornerstone of any management strategy for this fishery.

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to Norway lobster (Nephropsnorvegicus) by functional unit in ICES Subarea 7 and also demersal stocks.

### 18.7 Quality of assessment and forecast

Biological sampling for this stock is adequate. Since 2002, a dedicated annual UWTV survey has provided abundance estimates for the Aran Grounds with high precision. The area of the Aran Grounds was revised in 2015, resulting in a recalculation of the abundance time-series, which now also includes Galway Bay and Slyne Head. A number of other biological parameters such as mean weights and length distributions have also been revised. The revisions were made as part of an inter-benchmark process and have improved the quality of the assessment.
In the provision of catch options based on the absolute survey estimates, additional uncertainties related to mean weight in the landings and the discard rates also arise. From 2016, fisheries catching Nephrops in Subarea 7 are covered by the EU landings obligation (EU, 2015). Creel fisheries are exempted from the landings obligation, with a de minimis exemption consisting of a $5 \%$ discard rate by weight for the trawl fishery in 2019 (reduced from $6 \%$ in 2018 and $7 \%$ in both 2016 and 2017). The average discard rate by weight for FU17 over the last three years is $12.4 \%$. Catch advice and scenarios are provided this year on the assumption that discarding is assumed to continue at recent average.

Irish discard survival experiments indicate that the trawl discard survival may be around $64 \%$ (BIM, 2017). As a result, an exemption from the landings obligation based on high survivability has been granted by the European Commission. ICES continues to use the survival rate of $25 \%$ (ICES, 2016) as the survival rates estimated by BIM (2017) have not been evaluated by ICES.

There are several key uncertainties andbias sources in the method used here (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007;WKNEPHBID 2008; SGNEPS 2009; WGNEPS 2014; WKNEPS 2016). Ultimately, there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that is more accurate, although no more precise WKNEPH (ICES, 2009).

Landings data were adjusted to take into account landings that hadbeen misreported from FU16 from 2011 to 2017. This adjustment is thought to be reasonably accurate (See Section 18).

### 18.8 Recommendation for next benchmark

This stock was last benchmarked by IBPNeph (ICES, 2015a). WGCSE will keep the stock under close review and recommend future benchmark as required.

### 18.9 Management considerations

A meeting was held with stakeholders in March 2015 to discuss the state of the Aran Nephrops stock. In response to this meeting, voluntary effort limits were put in place for April, May and June 2015. These voluntary measures have significantly reduced effort and catches on the Aran grounds in 2015 before the UWTV survey.

Small whole Nephrops are the main species comprising the discards. The main fish species discarded are haddock, hake, whiting, megrim and dogfish (Anon, 2011).

The ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level. Management at the Functional Unit level could provide controls to ensure effort and catch were in line with resources available.

### 18.10 References

Anon. 2011. Atlas of Demersal Discarding, Scientific Observations and Potential Solutions, Marine Institute, Bord Iascaigh Mhara, September 2011. ISBN 978-1-902895-50-5. 82 pp.

Aristegui, M., Doyle, J., O’Brien, S., Fitzgerald, R., Vacherot, J.P., Sugrue, S. and Quinn, M. 2019. Aran, Galway Bay and Slyne Head Nephrops Grounds (FU17) 2019 UWTV Survey Report and catch scenarios for 2020. Marine Institute UWTV Survey report. In draft.

BIM. 2014. Catch comparison of Quad and Twin-rig trawls in the Celtic Sea Nephrops fishery. Bord Iascaigh Mhara, June 2014. http://www.bim.ie/media/bim/content/publications/Catch,compari-son,of,Quad,and,Twin-rig,trawls,in,the,Celtic,Sea,Nephrops,fishery,Trial,Report,2014.pdf
BIM. 2017. Report on Nephrops survivability in the Irish demersal trawl fishery. Bord Iascaigh Mhara, 29 September 2017. http://www.bim.ie/media/bim/content/publications/fisheries/6882-BIM-nephrops-survival-report-final.pdf

EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. Official Journal of the European Union, L 83: 1-17. http://data.europa.eu/eli/reg/2019/472/oj.

Marrs S.J., Atkinson R.J.A., Smith C.J., Hills J.M. 1996. Calibration of the towed underwater TV technique for use in stock assessment of Nephrops norvegicus. EC DGXIV Final Report, Study Project 94/069.

ICES. 2007. Report of the Workshop on the use of UWTV surveys for determining abundance in Nephrops stocks throughout European waters (WKNEPHTV). ICES CM: 2007/ACFM:14.
ICES. 2008. Report of the Workshop and training course on Nephrops burrow identification (WKNEPHBID). ICES CM 2008/LRC:03.

ICES. 2009. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 2009/LRC: 15, pp 52.
ICES. 2009. Report of the Benchmark Workshop on Nephrops assessment (WKNEPH). ICES CM 2009/ACOM:33.
ICES. 2012. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 2012/SSGESST:19.
ICES. 2014. Report of the Working Group on Nephrops Surveys (WGNEPS). ICES CM 2014/SSGESST:20. 57 pp .

ICES. 2015a. Report of the Inter-Benchmark Protocol of Nephrops in FU 17 and 14 (IBPNeph). ICES CM 2015/ACOM:38.

ICES. 2015. Report of the Working Group on Nephrops Surveys (WGNEPS), 10-13 November 2015, Cadiz, Spain. ICES CM 2015/ SSGIEOM:30.52 pp.

ICES. 2015. Report of the International Bottom Trawl Survey Working Group (IBTSWG), 23-27 March 2015, Bergen, Norway. ICES CM 2015/ SSGIEOM:20. 124 pp.
ICES. 2016. Report of the Working Group on Nephrops Surveys (WGNEPS), 7-8 November 2016, Reykjavík, Iceland. ICES CM 2016/ SSGIEOM:33.67 pp.
ICES. 2016. Report of the Workshop on Nephrops Burrow counting (WKNEPS), 9-11 November 2016, Reykjavík, Iceland. ICES CM 2016/ SSGIEOM:34.65 pp.

ICES. 2016. EU request to ICES to provide Fmsy ranges for selected stocks in ICES subareas 5 to 10. In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.2.3.1.

ICES. 2016a. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 183 pp.

Table 21.2.1. Nephrops in FU17 (Aran Grounds). Landings in tonnes by country.

| Year | France | Rep. of Ireland | UK | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1974 | 477 |  |  | 477 |
| 1975 | 822 |  |  | 822 |
| 1976 | 131 |  |  | 131 |
| 1977 | 272 |  |  | 272 |
| 1978 | 481 |  |  | 481 |
| 1979 | 452 |  |  | 452 |
| 1980 | 442 |  |  | 442 |
| 1981 | 414 |  |  | 414 |
| 1982 | 210 |  |  | 210 |
| 1983 | 131 |  |  | 131 |
| 1984 | 324 |  |  | 324 |
| 1985 | 207 |  |  | 207 |
| 1986 | 147 |  | 1 | 148 |
| 1987 | 62 |  | 0 | 62 |
| 1988 | 14 | 814 |  | 828 |
| 1989 | 27 | 317 | 3 | 347 |
| 1990 | 30 | 489 |  | 519 |
| 1991 | 11 | 399 |  | 410 |
| 1992 | 11 | 361 | 2 | 374 |
| 1993 | 11 | 361 | 0 | 372 |
| 1994 | 18 | 707 | 4 | 729 |
| 1995 | 91 | 774 | 2 | 867 |
| 1996 | 2 | 519 | 7 | 528 |
| 1997 | 2 | 839 | 0 | 841 |
| 1998 | 9 | 1401 | 0 | 1410 |
| 1999 | 0 | 1140 | 0 | 1140 |
| 2000 | 1 | 879 | 0 | 880 |
| 2001 | 1 | 912 | 0 | 913 |


| Year | France | Rep. of Ireland | UK | Total |
| :---: | :---: | :---: | :---: | :---: |
| 2002 | 2 | 1152 | 0 | 1154 |
| 2003 | 0 | 933 | 0 | 933 |
| 2004 | 0 | 525 | 0 | 525 |
| 2005 | 0 | 778 | 0 | 778 |
| 2006 | 0 | 637 | 0 | 637 |
| 2007 | 0 | 913 | 0 | 913 |
| 2008 | 0 | 1050 | 7 | 1057 |
| 2009 | 0 | 625 | 0 | 625 |
| 2010 | 0 | 930 | 9 | 939 |
| 2011 | 0 | 659 | 0 | 659 |
| 2012 | 0 | 1246 | 0 | 1246 |
| 2013 | 0 | 1295 | 0 | 1295 |
| 2014 | 0 | 766 | 0 | 766 |
| 2015 | 0 | 370 | 0 | 370 |
| 2016 | 0 | 641 | 0 | 641 |
| 2017 | 0 | 295 | 0.4 | 295 |
| 2018 | 0 | 494 | 42 | 536 |

Table 21.2.2. Nephrops in FU17 (Aran Grounds). Effort data for the Irish otter trawl Nephrops directed fleet.

| Year | Effort (Kw Days) | Landings (Kgs) |
| :---: | :---: | :---: |
| 1995 | 286,939 | 522,007 |
| 1996 | 174,030 | 312,421 |
| 1997 | 260,676 | 442,218 |
| 1998 | 445,308 | 940,902 |
| 1999 | 366,839 | 782,407 |
| 2000 | 293,684 | 561,244 |
| 2001 | 362,754 | 586,462 |
| 2002 | 350,346 | 798,744 |
| 2003 | 492,284 | 801,813 |
| 2004 | 355,673 | 420,652 |
| 2005 | 396,202 | 708,540 |
| 2006 | 337,503 | 618,515 |
| 2007 | 460,396 | 905,282 |
| 2008 | 512,245 | 1,052,077 |
| 2009 | 319,873 | 613,220 |
| 2010 | 441,080 | 910,346 |
| 2011 | 332,300 | 667,564 |
| 2012 | 488,721 | 1,139,413 |
| 2013 | 571,916 | 1,239,469 |
| 2014 | 460,818 | 774,097 |
| 2015 | 232,190 | 461,409 |
| 2016 | 396,502 | 578,420 |
| 2017 | 277,117 | 258,052 |
| 2018 | 233,793 | 483,723 |

Table 21.2.3. Nephrops in FU17 (Aran Grounds). Sampling levels.

| Year | Quarter | Number of samples |  | Numbers Measured |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Catch | Discards |
| 2008 | 1 | 2 | 3 | 565 | 1376 |
| 2008 | 2 | 9 | 8 | 2224 | 3758 |
| 2008 | 3 | 5 | 4 | 1266 | 1834 |
| 2008 | 4 | 3 | 3 | 889 | 1733 |
| 2009 | 1 | 3 | 3 | 800 | 1184 |
| 2009 | 2 | 6 | 6 | 1685 | 1978 |
| 2009 | 3 | 6 | 6 | 2260 | 2726 |
| 2009 | 4 | 2 | 2 | 1491 | 1149 |
| 2010 | 1 | 4 | 4 | 3322 | 2322 |
| 2010 | 2 | 8 | 7 | 3577 | 2957 |
| 2010 | 3 | 2 | 2 | 951 | 742 |
| 2010 | 4 | 6 | 4 | 3209 | 1802 |
| 2011 | 1 | 7 | 7 | 3755 | 3537 |
| 2011 | 2 | 7 | 7 | 7399 | 6617 |
| 2011 | 3 | 4 | 2 | 3531 | 2386 |
| 2011 | 4 | 5 | 5 | 2440 | 2271 |
| 2012 | 1 | 3 | 3 | 1538 | 1250 |
| 2012 | 2 | 17 | 15 | 6481 | 5113 |
| 2012 | 3 | 0 | 0 | - | - |
| 2012 | 4 | 5 | 5 | 2333 | 1945 |
| 2013 | 1 | 10 | 9 | 3108 | 2983 |
| 2013 | 2 | 11 | 11 | 3733 | 3733 |
| 2013 | 2 | 3 | 3 | 1163 | 1263 |
| 2013 | 4 | 7 | 7 | 2956 | 1779 |
| 2014 | 1 | 3 | 3 | 1208 | 1223 |
| 2014 | 2 | 12 | 12 | 5365 | 3563 |
| 2014 | 3 | 2 | 2 | 786 | 499 |
| 2014 | 4 | 8 | 8 | 3542 | 2760 |


| Year | Quarter | Number of samples |  | Numbers Measured |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Catch | Discards |
| 2015 | 1 | 2 | 2 | 827 | 611 |
| 2015 | 2 | 2 | 2 | 961 | 664 |
| 2015 | 3 | 0 | 0 | - | - |
| 2015 | 4 | 2 | 2 | 1047 | 1388 |
| 2016 | 1 | 5 | 4 | 2292 | 876 |
| 2016 | 2 | 11 | 11 | 4756 | 3383 |
| 2016 | 3 | 6 | 5 | 3020 | 2048 |
| 2016 | 4 | 6 | 6 | 1389 | 1311 |
| 2017 | 1 | 3 | 3 | 1214 | 845 |
| 2017 | 2 | 6 | 4 | 2911 | 1569 |
| 2017 | 3 | 2 | 1 | 1018 | 223 |
| 2017 | 4 | 3 | 3 | 1176 | 839 |
| 2018 | 1 | 3 | 3 | 1224 | 1241 |
| 2018 | 2 | 8 | 8 | 3179 | 2971 |
| 2018 | 3 | 1 | 1 | 467 | 388 |
| 2018 | 4 | 6 | 6 | 1894 | 2487 |

Table 21.2.4. Nephrops in FU17 (Aran Grounds). Raised landings and discard weight and numbers by year.

| Year | Landings (t) | Discards (t) | Landings in number ('000s) | Discards in number ('000s) | Discards by weight (\%) | Discards by number (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 1057 | 248 | 48,162 | 22,074 | 19.0 | 31.4 |
| 2009 | 626 | 129 | 24,935 | 9,487 | 17.1 | 27.6 |
| 2010 | 939 | 224 | 37,341 | 15,246 | 19.3 | 29.0 |
| 2011 | 659 | 92 | 31,950 | 8,542 | 12.2 | 21.1 |
| 2012 | 1246 | 86 | 61,076 | 8,292 | 6.5 | 12.0 |
| 2013 | 1295 | 129 | 60,016 | 12,034 | 9.1 | 16.7 |
| 2014 | 766 | 48 | 33,882 | 5,038 | 5.9 | 12.9 |
| 2015 | 370 | 15 | 17,693 | 1,622 | 3.8 | 8.4 |
| 2016 | 641 | 69 | 30,231 | 6,375 | 9.7 | 17.4 |
| 2017 | 295 | 38 | 13,269 | 3,605 | 11.3 | 21.4 |
| 2018 | 536 | 106 | 22,049 | 10,490 | 16.5 | 32.2 |

Table 21.2.5. Nephrops in FU17 (Aran Grounds). Results summary table for geostatistical analysis of UWTV survey.

| Ground | Year | Number of stations | Mean Density adjusted (burrow $/ \mathrm{m}^{2}$ ) | Domain <br> Area (km²) | Geostatistical Abundance Estimate adjusted (millions burrows) | CV on Burrow estimate \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Aran Grounds | 2002 | 49 | 0.79 | 1196 | 947 | 3 |
|  | 2003 | 41 | 0.94 | 1196 | 1118 | 6 |
|  | 2004 | 64 | 1.08 | 1196 | 1297 | 3 |
|  | 2005 | 70 | 0.81 | 1196 | 972 | 2 |
|  | 2006 | 67 | 0.46 | 1196 | 556 | 3 |
|  | 2007 | 71 | 0.69 | 1196 | 828 | 2 |
|  | 2008 | 63 | 0.41 | 1196 | 494 | 3 |
|  | 2009 | 82 | 0.52 | 1196 | 627 | 2 |
|  | 2010 | 87 | 0.63 | 1196 | 752 | 2 |
|  | 2011 | 76 | 0.51 | 1196 | 609 | 2 |
|  | 2012 | 31* | 0.33 | 1196 | 397 | 3 |
|  | 2013 | 31* | 0.33 | 1196 | 390 | 4 |
|  | 2014 | 33* | 0.28 | 1196 | 332 | 4 |
|  | 2015 | 34* | 0.40 | 1197 | 480 | 4 |
|  | 2016 | 34* | 0.29 | 1197 | 343 | 3 |
|  | 2017 | 31* | 0.31 | 1196 | 377 | 3 |
|  | 2018 | 33* | 0.40 | 1196 | 488 | 3 |
|  | 2019 | 31* | 0.38 | 1196 | 458 | 4 |

[^11]Table 21.2.6. Nephrops in FU17 (Galway Bay and Slyne Head). Results summary table for analysis of UWTV survey. Random stratified estimates given for these grounds only.

| Ground | Year | Number of stations | Mean Density adjusted (burrow/m²) | Domain Area (km²) | Raised Abundance Estimate adjusted (millions burrows)* | CV on Burrow estimate \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Galway Bay | 2002 | 7 | 1.18 | 79.0 | 93.1 | 7 |
|  | 2003 | 3 | 1.30 | 79.0 | 102.6 | 16 |
|  | 2004 | 8 | 1.17 | 79.0 | 92.2 | 14 |
|  | 2005 | 4 | 1.30 | 79.0 | 103.0 | 11 |
|  | 2006 | 3 | 0.74 | 79.0 | 58.8 | 9 |
|  | 2007 | 5 | 0.91 | 79.0 | 71.8 | 8 |
|  | 2008 | 5 | 0.40 | 79.0 | 31.6 | 4 |
|  | 2009 | 8 | 0.71 | 79.0 | 56.3 | 4 |
|  | 2010 | 10 | 1.24 | 79.0 | 97.6 | 11 |
|  | 2011 | 6 | 0.55 | 79.0 | 43.2 | 12 |
|  | 2012 | 4 | 0.64 | 79.0 | 50.9 | 10 |
|  | 2013 | 5 | 0.37 | 79.0 | 29.6 | 10 |
|  | 2014 | 3 | 0.50 | 79.0 | 39.8 | 6 |
|  | 2015 | 5 | 0.71 | 79.0 | 55.8 | 15 |
|  | 2016 | 7 | 0.32 | 79.0 | 25.1 | 7 |
|  | 2017 | 5 | 0.20 | 79.0 | 15.8 | 4 |
|  | 2018 | 5 | 0.41 | 79.0 | 32.5 | 17 |
|  | 2019 | 5 | 0.29 | 79.0 | 22.8 | 11 |
| Slyne Head | 2002 | 5 | 0.76 | 39.1 | 29.8 | 8 |
|  | 2003* | 0 | 0.65 | 39.1 | 25.3 | 0 |
|  | 2004 | 3 | 0.53 | 39.1 | 20.8 | 10 |
|  | 2005 | 3 | 0.44 | 39.1 | 17.4 | 1 |
|  | 2006 | 3 | 0.30 | 39.1 | 11.8 | 9 |
|  | 2007 | 4 | 0.51 | 39.1 | 19.8 | 12 |
|  | 2008* | 0 | 0.41 | 39.1 | 16.0 | 0 |
|  | 2009 | 6 | 0.31 | 39.1 | 12.2 | 7 |
|  | 2010 | 7 | 0.73 | 39.1 | 28.7 | 4 |


| Ground | Year | Number of stations | Mean Density adjusted (burrow/m²) | Domain Area (km ${ }^{\mathbf{2}}$ ) | Raised Abundance Estimate adjusted (millions burrows)* | CV on Burrow estimate \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2011 | 7 | 0.51 | 39.1 | 20.0 | 5 |
|  | 2012 | 3 | 0.52 | 39.1 | 20.5 | 2 |
|  | 2013 | 4 | 0.54 | 39.1 | 21.1 | 10 |
|  | 2014 | 4 | 0.28 | 39.1 | 11.0 | 6 |
|  | 2015 | 5 | 0.50 | 39.1 | 19.6 | 4 |
|  | 2016 | 4 | 0.27 | 39.1 | 10.8 | 3 |
|  | 2017 | 4 | 0.27 | 39.1 | 10.7 | 4 |
|  | 2018 | 5 | 0.84 | 39.1 | 33.0 | 12 |
|  | 2019 | 5 | 0.29 | 39.1 | 11.5 | 8 |

*estimated as no survey data available for these years.

Table 21.2.7. Nephrops in FU17. Results summary table for analysis of UWTV survey for the combined grounds.

| Year | Abundance <br> (Millions) | Upper bound | Lower bound |
| :---: | :---: | :---: | :---: |
| 2002 | 1070 | 1154 | 985 |
| 2003 | 1246 | 1434 | 1059 |
| 2004 | 1410 | 1517 | 1302 |
| 2005 | 1092 | 1154 | 1030 |
| 2006 | 627 | 703 | 551 |
| 2007 | 920 | 982 | 858 |
| 2008 | 541 | 588 | 494 |
| 2009 | 696 | 739 | 653 |
| 2010 | 879 | 926 | 831 |
| 2011 | 672 | 720 | 624 |
| 2012 | 468 | 520 | 417 |
| 2013 | 441 | 506 | 376 |
| 2014 | 383 | 440 | 327 |
| 2015 | 556 | 627 | 484 |
| 2016 | 379 | 420 | 339 |
| 2017 | 404 | 445 | 362 |
| 2018 | 554 | 637 | 471 |
| 2019 | 493 | 558 | 427 |

Table 21.3.1. Nephrops in FU17 (Aran Grounds). Forecast inputs (bold) and historical estimates of mean weight in landings and harvest rate. Removals estimated in years with no sampling (*) using ratio of removals to landings in adjacent years. na= not available due to non-cooperation with sampling programmes.

|  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { 亠ু } \\ & \stackrel{\text { ® }}{2} \end{aligned}$ | millions | millions | millions | \% | \% | millions |  | \% | tonnes | tonnes | gramme | gramme |
| 2001 | 48.7 | 25.4 | 67.8 | 28.2 | 34.3 |  |  |  | 912 |  |  |  |
| 2002 | 54.5 | 17.7 | 67.8 | 19.6 | 24.5 | 1070 | 84 | 6.30 | 1152 | 192 | 21.2 | 10.8 |
| 2003 | 44.1 | 18.3 | 57.8 | 23.7 | 29.3 | 1246 | 188 | 4.60 | 933 | 183 | 21.2 | 10 |
| 2004 | 29 | 11.4 | 37.6 | 22.9 | 28.2 | 1410 | 107 | 2.70 | 525 | 112 | 18.1 | 9.9 |
| 2005 | 42.4 | 19.7 | 57.2 | 25.9 | 31.7 | 1092 | 62 | 5.20 | 778 | 182 | 18.4 | 9.2 |
| 2006 | na | na | 49.5* | na | na | 627 | 76 | 7.90 | 636 | na | na | na |
| 2007 | na | na | 57.3* | na | na | 920 | 62 | 6.20 | 913 | na | na | na |
| 2008 | 48.2 | 22.1 | 64.7 | 25.6 | 31.4 | 541 | 47 | 12.00 | 1057 | 248 | 21.9 | 11.2 |
| 2009 | 24.9 | 9.5 | 32 | 22.2 | 27.6 | 696 | 43 | 4.60 | 626 | 129 | 25.1 | 13.6 |
| 2010 | 37.3 | 15.2 | 48.8 | 23.4 | 29.0 | 879 | 47 | 5.60 | 939 | 224 | 25.2 | 14.7 |
| 2011 | 31.9 | 8.5 | 38.4 | 16.7 | 21.1 | 672 | 48 | 5.70 | 659 | 92 | 20.6 | 10.8 |
| 2012 | 61.1 | 8.3 | 67.3 | 9.2 | 12.0 | 468 | 52 | 14.40 | 1246 | 86 | 20.4 | 10.4 |


|  |  |  |  |  |  |  |  |  | $\begin{aligned} & \text { a0 } \\ & \text { =0 } \\ & \text { 듲 } \end{aligned}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { 末 } \\ & \stackrel{\text { ® }}{2} \end{aligned}$ | millions | millions | millions | \% | \% | millions |  | \% | tonnes | tonnes | gramme | gramme |
| 2013 | 60 | 12 | 69 | 13.1 | 16.7 | 441 | 65 | 15.70 | 1295 | 129 | 21.6 | 10.7 |
| 2014 | 33.9 | 5 | 37.7 | 10.0 | 12.9 | 383 | 57 | 9.80 | 766 | 48 | 22.6 | 9.6 |
| 2015 | 17.7 | 1.6 | 18.9 | 6.4 | 8.4 | 556 | 71 | 3.40 | 370 | 15 | 20.9 | 9.1 |
| 2016 | 30.2 | 6.4 | 35.0 | 13.7 | 17.4 | 379 | 41 | 9.20 | 641 | 69 | 21.2 | 10.9 |
| 2017 | 13.3 | 3.6 | 16.0 | 16.9 | 21.4 | 404 | 41 | 4.0 | 295 | 38 | 22.2 | 10.5 |
| 2018 | 22.0 | 10.5 | 29.9 | 26.3 | 32.2 | 554 | 83 | 5.4 | 536 | 106 | 24.3 | 10.1 |
| 2019 |  |  |  |  |  | 493 | 66 |  |  |  |  |  |
| Average 2016-2018 |  |  |  | 18.9 | 23.7 |  |  |  | Average 2008-2018 |  | 22.4 | 11.0 |



Figure 21.1.1. Nephrops in FU17 (Aran Grounds). Time-series of the number of Irish vessels reporting landings of Nephrops from FU17 with a $>10 \mathrm{t}$ threshold.


Figure 21.1.2. Nephrops in FU17 (Aran Grounds). Combined box and kite plot of Irish vessel's power on the Aran Grounds by year. The blue line indicates the mean.


Figure 21.2.1. Nephrops in FU17 (Aran Grounds). Landings in tonnes by country.


Figure 21.2.2. Nephrops in FU17 (Aran Grounds). Effort data (kW days) for Irish directed Nephrops fleet.


Figure 21.2.3. Nephrops FU17 (Aran Grounds). Sampling levels for the Aran grounds.

Length frequencies for catch (dotted) and landed(solid): Nephrops in FU17


Figure 21.2.4. Nephrops FU17 Aran Grounds. Annual length composition of catches (dotted line) and landings (solid line) for females (left) and males (right) from 2008 (bottom) to 2018 (top). Annual mean length of catches (dotted vertical line) and landings (solid vertical line) are also shown. Minimum Landing Size ( 25 mm ) and 35 mm levels are also displayed with vertical lines.


Figure 21.2.5. Nephrops FU17 (Aran Grounds). Proportion of males by number in the catch (blue) and landings (black).


Figure 21.2.6. Nephrops FU17 (Aran Grounds). Mean weight in catch samples by sex showing cyclical trends.


Figure 21.2.7. Nephrops FU17 (Aran Grounds). Annual mean weight (g) estimates of landings (blue) and discards (black).


Figure 21.2.8. Nephrops in FU17 (Aran Grounds). Contour plots of the krigged density estimates for the Aran Ground UWTV surveys from 2002 (top left) to 2018 (bottom centre).


Figure 21.2.9. Nephrops FU17 Aran Grounds. Nephrops burrow estimates in FU17 Aran (blue), Galway Bay (green) and Slyne Head (red) grounds 2002-2018.


Figure 21.2.10. Time-series of total abundance estimates for FU17 (error bars indicate $95 \%$ confidence intervals) and $B_{\text {trigger }}$ is dashed green line.


Figure 21.3.1. Nephrops FU17 Aran Grounds. Harvest Rate (\% dead removed/UWTV abundance). The dashed and solid lines are the MSY proxy and the harvest rate respectively.

# 19 Norway lobster (Nephrops norvegicus) in divisions 7.a, 7.g and 7.j, Functional Unit 19 (Irish Sea, Celtic Sea, eastern part of southwest of Ireland) 

## Type of assessment in 2019

This stock was benchmarked in February 2014 and the assessment and provision of catch advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the benchmark WG (ICES, 2014) and set out in the stock annex.

## ICES advice applicable to 2018

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2014-2016, catches in 2018 should be no more than 1192 tonnes.

To ensure that the stock in functional unit (FU) 19 is exploited sustainably, management should be implemented at the functional unit level."

## ICES advice applicable to 2019

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2015-2017, catches in 2019 should be no more than 173 tonnes.

To ensure that the stock in Functional Unit 19 is exploited sustainably, management should be implemented at the functional unit level."

### 19.1 General

## Stock description and management units

In FU19 Nephrops are caught on a large number of spatially discrete small inshore grounds and on some larger grounds further offshore and of these the 'Galley ground 4' and around Cork channels appear to be the most important (see Figure 22.1.1). The Nephrops stock (FU19) covers ICES rectangles;31-33 D9-E0;31E1;32E1-E2;33E2-E3 within 7.a, 7.g, and 7.j. This stock is included as part of the TAC Area 7 Nephrops, which includes the following stocks: Irish Sea East and West (FU14, FU15), Porcupine Bank (FU16), northwestern Irish Coast (FU18) and the Celtic Sea (FU20-22).

The map below shows FU19 assessment area (blue) and TAC area (red). There is no evidence that the individual functional units belong to the same stock. See Section 18 for details on Nephrops in Subarea 7 general section.


## Ecosystem aspects

This section is detailed in stock annex. There are no updates.

## Fishery description

A description of the fleet is given in the stock annex.
The time-series of numbers of vessels reporting landings greater than 10 t is updated in Figure 22.1.2. The numbers of vessels has been relatively stable since 1995 except in 2018 where there was a sharp decrease. The time-series of vessel power is shown as a box and kite plot in Figure 22.1.3.

## Fishery in 2018

There has been a trend for Irish vessels ( $>18 \mathrm{~m}$ ) to switch to multi (quad) rig trawls. Provisional data suggest a $\sim 30 \%$ increase in Nephrops catch rates and a reduction in fish bycatch of $\sim 30 \%$ due to the lower headline height. The number of French vessels reporting landings in FU19 has decreased from 35 vessels in 2005 to seven vessels in 2018.

## Information from stakeholders

None available.

### 19.2 Data

## InterCatch

All data were available in InterCatch and used for catch data only. French catch data provided directly by the national expert and not extracted from InterCatch.

## Landings

Landings data for FU19 are summarized in Table 22.2.1. The Republic of Ireland, France and the UK report landings for FU19. Landings data for Ireland were revised back to 2008, which resulted in minor revisions in the order of 1 to $5 \%$ (stock annex). These revised data have been used in the assessment this year. The Republic of Ireland landings have fluctuated considerably throughout the time-series, with a marked dip in 1994 (Table 22.2.1; Figure 22.2.1). The highest landings in the time-series were observed in 2002-2004 (>1000 t). Landings in 2005 and 2006 have been below average for the series. In 2017, landings decreased by approximately $30 \%$ for the Irish fleet and were below the series average. This can beexplained due to the poor weather conditions in quarter 1, which hampered fishing activities of smaller vessels and the larger vessels maximising effort in other FUs. In 2018, landings further decreased by $50 \%$ compared to 2017. Landings by the French fleet have fluctuated with a declining trend throughout the time-series from the highest value in 1989 of 245 t to 4 t in 2018. Landings from the UK are minor at 4 t .

## Effort

In line with WGCSE 2015 recommendation effort is reported in KWdays and lpue reported in KG/kwdays in the knowledge that the trend is likely to be a biased underestimate because it is not adjusted for efficiency or behavioural changes. The effort series is based on the same criteria for FU15, 16, 17, 22 and 20-21 ( $30 \%$ landings threshold) and will be contingent on the accuracy of landings data reported in logbooks.

Disaggregated effort and landings data are available for the Irish Nephrops directed fleet in FU19 from 1995-2018 for all vessels and vessels $>18$ metres total length. (Table 22.2.2; Figure 22.2.2). For vessels $>18$ effort (sinceearly 2000s) has fluctuated with an overall decreasing trend in recent three years. This can be explained by fleet mobility where vessels target Nephrops in this area in periods of good emergence. For vessels $<18$ effort has decreased in 2017 to 2018 due to weather conditions.

## Sampling levels

Sampling levels, data aggregating and raising procedures were reviewed by WKCELT 2014, and are documented in the stock annex. The time-series of samples is shown in Figure 22.2.3 and Table 22.2.3. Sampling levels in 2018 were good and are comparable to 2017 levels.

## Commercial length-frequency distributions

Spatial and temporal coverage is problematic with landings from FU19 coming from several discrete grounds (see stock annex.) The sampling intensity and coverage has varied over the timeseries (see stock annex). Since 2008, samplinghas been good although the majority of the samples come from Bantry Bay recently. Also, sampling of the discards is quite sparse over the time-series and are difficult to obtain due to the spatial coverage of the grounds. The catch samples from 2008 to 2018 were split using the discard selection ogive agreed at the benchmark. The lengthweight regression parameters given in the stock annex are used to calculate sampled weights and appropriate quarterly raising factors. The length distributions are shown in Figure 22.2.4. The mean size has remained relatively stable and the trend in mean size is stable in recent years.

## Sex ratio

The sex ratio in the landings is male biased in most years but there is a trend towards increased percentage of females in the landings (Figure 22.2.5). The proportion of females was higher in 2013 and this was confirmed by the industry.

## Mean weight explorations

Explorations of the mean weight in the catch samples by sex shows a strong cyclical pattern in the females for Bantry Bay (Figure 22.2.6) and also, all grounds combined (Figure 22.2.7). This corresponds with the emergence of mature females from the burrows to mate in summer. These data also show an increase in mean weights for males in 2016. The annual mean weight estimate for landings and discards is shown in Figure 22.2.8. The landings mean weight estimates show a slight decrease in 2018.

## Discarding

Sampling of the discards has quite sparse over the time-series and are difficult to obtain due to the spatial coverage of the grounds (see stock annex). Since 2002, discard rates have been estimated using unsorted catch and discards sampling (as described in the stock annex). WKCELT 2014 examined the available discard data observations for FU19. An average discard selection ogive using data from Bantry Bay in years 2008 and 2013 was generated and deemed appropriate given the variable sampling intensity and coverage. The catch data from 2008 to 2013 were then revised and split into landings and discards. Catch data sampling for years previous to 2008 were not revised as was considered to be not of good enough quality. The 2018 catch data were split using this selection ogive since 2008.
Discard rates range between $25-86 \%$ of total catch by weight and $40-80 \%$ of total catch by number (Table 22.2.4). These high discard rates are very high compared with other FUs. This is because the fleet is mainly smaller inshore vessels with limited space for extra crew. On-board "tailing" of the smaller Nephrops is not usually practised and the bigger Nephrops are picked from catches. There is no information on discard survival rate in this fisherybut a 25\% discard survival rate is assumed in line with other Nephrops stocks in the Celtic Sea.

Gear selectivity trials by Bord Iascaigh Mhara (BIM, 2017) reported a $64 \%$ survivor rate for Nephrops caught in a trawl with a SELTRA selectivity device in the outer Galway Bay area.

Table 22.3.1 gives weights, numbers and mean weights of the landings and discard raised internationally according to the stock annex.

## Abundance indices from UWTV surveys

The methods used during the survey were similar to those employed for UWTV surveys of Nephrops stocks around Ireland and elsewhere are documented by WKNEPHTV (ICES, 2007), WKNEPHBID (ICES, 2008), SGNEPS (ICES, 2009; 2010; 2012), WGNEPS (ICES, 2013; 2014; 2015; 2016a; 2017; 2018a), WKNEPS (ICES, 2016b; 2018b) and Leocádio, A.et al., 2018. Given the scale of the area and the number of distinct patches, it is unrealistic to expect sufficient stations ( $\sim 10$ ) in each individual patch to estimate densities separately. The random stratified approach may cause problems in years where the planned survey coverage is not achieved. WKCELT 2014 concluded that WGCSE or WGNEPS should make recommendations on the most appropriate fill in procedure to be adopted in these cases.

The spatial extent of the Nephrops grounds in FU19 has been re-defined by WKCELT 2014 and the abundance estimates are calculated using these areas. The redefinition of the polygons in FU19 resulted in $\sim 16 \%$ increase in overall area from $1653 \mathrm{~km}^{2}$ to $1973 \mathrm{~km}^{2}$ (see stock annex). The discrete grounds have been named as: Bantry Bay, Galley Ground 1-4, Cork Channels and Helvick 1-2 and are shown in Figure 22.1.1. In terms of area the Galley Grounds (1-4) account for $61 \%$ of the total grounds in FU19 and Galley Ground 4 is the largest of these representing $47 \%$ of the total area (Table 22.2.5). Helvick patches 2 and 3 were also amalgamated and renamed Helvick 2 based on the information from the VMS data.

From 2011 to 2019, an average of 40 stations have been completed annually. The survey design is based on randomly picked stations from the ground polygons and the sampling effort on each ground was determined by relative area.
All grounds except Galley Ground 4 in 2011 and Galley Ground 1 in 2012 were covered by the TV survey. Since 2015, a new patch Kenmare Bay was surveyed. Operational details of the 2019 UWTV survey are available (Doyle, 2019).
Detailed summary statistics for the various Nephrops patches in FU19 over the time-series are presented in Table 22.2.6. The mean density varies across the different patches, but there is some consistency to the estimates over time. The UWTV coverage has improved. In 2019, all discrete grounds were covered by the TV survey and also two stations on a new patch Dunmanus Bay (Doyle et al., 2019).

The 2019 mean density estimates vary between patches from the lowest value 0.0 (no. $/ \mathrm{m}^{2}$ ) observed at Helvick 2 to the highest observed at 0.66 (no. $/ \mathrm{m}^{2}$ ) at Galley ground 2 (Table 22.2.6, Figure 22.2.9). The overall mean density for FU19 in 2019 is 0.20 (no. $/ \mathrm{m}^{2}$ ) which is the second lowest observed in the time-series (Table 22.2.7).

Figure 22.2.10 and Table 22.2 .7 shows the total abundance estimate for FU19 with the WKMSYRef4 proposed MSY Btrigger (ICES, 2016XX, ICESYY). The 2019 abundance estimate was 220\% higher than in 2018 and at 386 million is below the MSY B trigger ( 430 million) with a RSE of 15\%, which is below the $20 \%$ limit recommended by SGNEPs (2012).

## Information from Irish Groundfish survey

Length-frequency data of the Nephrops catches on the Irish groundfish survey (IGFS-WIBTS-Q4) from 2003-2017 are available (Stokes et al., 2014; ICES, 2015). These data were investigated for trends in indicators such as possible recruitment signals (Figure 22.2.11). The mean size of males and females in from the survey was fairly stable over time at 33 mm for males and 25 mm for females.

### 19.3 Assessment

## Comparison with previous assessments

The WGCSE 2019 carried out an UWTV based assessment for this stock. The methods used were very much in line with WKNEPH (ICES, 2009) and the approach taken for other Nephrops stocks in 6 and 7 by WGCSE. This approach was benchmarked at WKCELT 2014 (ICES, 2014).

## State of the stock

UWTV abundance estimates suggest that the stock size has fluctuated although the series is quite short. The 2019 estimate is the second lowest observed and is below the MSY Btrigger. The 2019 abundance remains below the average of the series (geomean: [2011-2019]: 471 million).

Table 22.3.1 summarizes recent abundance estimates, harvest rates for the stock along with other stock parameters. Harvest rate is calculated as (landings + dead discards)/(abundance estimate).

Table 22.3.1.and Figure 22.3.1 summarize recent harvest ratios, which have been below the Fmsyproxy for the last three years.

### 19.4 Catch scenario table

Catch scenario table inputs and historical estimates of mean weight in landings and harvest ratios are presented in Table 22.3.1 and summarised below.

The basis for the catch options.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Stock abundance <br> viduals | UWTV survey 2019 |  |
| Mean weight in <br> wanted catch | 29.0 g | Average 2016-2018 |
| Mean weight in un- <br> wanted catch | 14.8 g | Average 2016-2018 |
| Unwanted catch <br> rate | $38.5 \%$ | Average 2016-2018 (by number). Calculated as discards divided by land- <br> ings + discards. |
| Discard survival rate | $25.0 \%$ | Only applies in scenarios where discarding is allowed. |
| Dead unwanted <br> catch rate | $31.9 \%$ | Average 2016-2018 (by number). Calculated as dead discards divided by <br> removals (landings + dead discards). Only applies in scenarios where dis- <br> carding is allowed. |

The average in the recent three years is used to calculate the mean weight for landings and discards. The discard rates and proportions for the last three years are used to account for recent on-board retention practices (this is also according to the stock annex).

A prediction of landings for the FU19 using the approach agreed procedure proposed at WKNEPH 2009 and outlined in the stock annex will be made on the basis of the 2018 UWTV survey. This will be presented in October 2019 for the provision of advice.

### 19.5 Reference points

WKMS YRef4 updated the Fmsy reference points for FU19 (ICES, 2016XX; 2016YY) on the basis of an average of estimated $\mathrm{Fmsy}^{\text {p }}$ proxy harvest rates over a period of years, this corresponds more closely to the methodology for finfish. The updated harvest rate calculated at $9.3 \%$ is expected to deliver high long-term yield with a low probability of recruitment overfishing. This is close to the harvest rate of $8.1 \%$ calculated by WKCELT (ICES, 2014)

This stock previously did not have MSY Btrigger specified, the time-series and range of indicator biomass is also limited such that direct use of Blossis considered too close to equilibrium biomass. The workshop proposed to use the $5 \%$ interval on the probability distribution of indicator biomass assuming a normal distribution, which is analogous to the $5 \%$ on BMSY proposed for finfish stocks assuming these Nephrops FU have been exploited at a rate close to near HRMSY. The MSY Btrigger for FU 19 is 434 million individuals rounded to 430 million.

These reference points shown in text table below should remain under review by WGCSE should improved data become available.

| Stock code | MSY Flower* | FMSY* | MSY Fupper* with AR | MSY Btrigger | MSY Fupper* with no AR |
| :--- | :---: | :---: | :---: | :---: | :---: |
| nep-19 | $8.3 \%$ | $9.3 \%$ | $9.3 \%$ | $430^{* * *}$ | $9.3 \%$ |

* Harvest rate (HR).
*** Abundance in millions.


### 19.6 Management strategies

No specific management plan exists for this stock.
The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to Norway lobster (Nephropsnorvegicus) by functional unit in ICES Subarea 7 and also demersal stocks.

### 19.7 Quality of assessment and forecast

Biological sampling for this stock is improving given the spatial distribution of the Nephrops mud patches. A number of other biological parameters such as mean weights and length distributions have also been revised. The revisions were made as part of the benchmark process and have improved the quality of the assessment.

In the provision of catch options based on the absolute survey estimates, additional uncertainties related to mean weight in the landings and the discard rates also arise. For FU19 deterministic estimates of the mean weight in the landings and discard rates for 2016-2018 are used, although there is some variability of these over time.

From 2016, fisheries catching Nephrops in Subarea 7 are covered by the EU landings obligation (EU, 2015). Creel fisheries are exempted from the landings obligation, with a de minimis exemption consisting of a $5 \%$ discard rate by weight for the trawl fishery in 2019 (reduced from $6 \%$ in 2018 and 7\% in both 2016 and 2017).

Irish discard survival experiments indicate that the trawl discard survival may be around $64 \%$ (BIM, 2017). As a result, an exemption from the landings obligation based on high survivability has been granted by the European Commission. The average discard rate by weight for FU19 over the last three years is $24 \%$. Catch advice and scenarios are provided this year on the assumption that discarding is assumed to continue at the recent average.

There are several key uncertainties and bias sources in the method used here (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007; WKNEPHBID 2008; SGNEPS 2009; WGNEPS 2014). Ultimately, there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that is more accurate, although no more precise WKNEPH (ICES, 2009).

Different densities are apparent on the various different grounds within this FU. For the 2019 survey, the number of observations on each individual patch is relatively low making the relative standard error (RSE) estimates not that relevant. Aggregating all areas together gives a mean burrow density of 0.20 with a RSE of around $17 \%$, which is below the $20 \%$ threshold recommended by SGNEPS (ICES, 2012). The cumulative bias estimates for FU19 are largely based on expert opinion. The precision of these bias corrections cannot yet be characterized, but is likely to be lower than that observed in the survey.

Landings data are adjusted to take into account landings that have been misreported from FU16 since 2011. This adjustment is thought to be reasonably accurate (See Section 19).

### 19.8 Recommendations for next benchmark

This stock was benchmarked by ICES in February 2014 (ICES, 2014). WGCSE will keep the stock under close review and recommend future benchmark as required.

### 19.9 Management considerations

The trends from the fishery (landings, effort, mean size, etc.) appear to be relatively stable. The UWTV abundance and mean density estimates vary between the discrete patches and population dynamics between these are not fully understood. The 2019 survey result is the second lowest observed in the time-series.

In recent years, several newer vessels specializing in Nephrops fishing have participated in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates. Since the introduction of effort management associated with the cod long-term plan(EC 1342/2008) therehave been concerns that effort will be displaced towards FU19 and other Nephrops grounds where effort control has not been put in place.
Nephrops fisheries in this area are fairly mixed, also catching megrim, anglerfish and other demersal species. There are also some catches of hake, and in the offshore parts of the area. The Nephrops grounds in FU19 coincide with an important nursery area for juvenile hake and anglerfish among other species (ICES, 2009).

### 19.10 References

Anon. 2011. Atlas of Demersal Discarding, Scientific Observations and Potential Solutions, Marine Institute, Bord Iascaigh Mhara, September 2011. ISBN 978-1-902895-50-5. 82 pp.

Charuau A., Morizur Y., Rivoalen J.J. 1982. Survival of discarded Nephrops norvegicus in the Bay of Biscay and in the Celtic Sea. ICES CM 1982/B:13.

EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. Official Journal of the European Union, L 83: 1-17. http://data.europa.eu/eli/reg/2019/472/oj.

Doyle, J., Aristegui, M., O’Brien, S., Lynch, D., Vacherot, JP. and Fitzgerald, R. 2019. FU19 Nephrops Grounds (FU19) 2019 UWTV Survey Report and catch scenarios for 2020. Marine Institute UWTV Survey report. in draft.

Gerritsen, H.D. and Lordan, C. 2011. Integrating Vessel Monitoring Systems (VMS) data with daily catch data from logbooks to explore the spatial distribution of catch and effort at high resolution. ICES Journal of Marine Science, 68(1): 245-453.

BIM. 2017. Report on Nephrops survivability in the Irish demersal trawl fishery. Bord Iascaigh Mhara, 29 September 2017 .http://www.bim.ie/media/bim/content/publications/fisheries/6882-BIM-nephrops-survival-report-final.pdf.
ICES. 2007. Report of the Workshop on the use of UWTV surveys for determining abundance in Nephrops stocks throughout European waters (WKNEPHTV). ICES CM: 2007/ACFM:14.
ICES. 2008. Report of the Workshop and training course on Nephrops burrow identification (WKNEPHBID). ICES CM 2008/LRC:03.

ICES. 2009. Report of the Benchmark Workshop on Nephrops assessment (WKNEPH). ICES CM 2009/ACOM:33.
ICES. 2010. Report of the Study Group on Nephrops Surveys (SGNEPS), 9-11 November 2010, Lisbon, Portugal. ICES CM 2010/SSGESST:22.
ICES. 2012. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 012/SSGESST:19.
ICES. 2013. Report of the Working Group on Nephrops Surveys (WGNEPS), 5-8 November 2013, Barcelona, Spain. ICES CM 2013/SSGESST:21. 27 pp.
ICES. 2014. Report of the Benchmark Workshop on Celtic Sea stocks (WKCELT), 3-7 February 2014, ICES Headquarters, Copenhagen, Denmark. ICES CM $2014 \backslash$ ACOM:42. 194 pp.

ICES. 2014. Report of the Working Group on Nephrops Surveys (WGNEPS), 4-6 November 2014, Lisbon, Portugal. ICES CM 2014/SSGESST:20.57 pp.
ICES. 2015. Report of the Working Group on Nephrops Surveys (WGNEPS), 10-13 November 2015, Cadiz, Spain. ICES CM 2015/ SSGIEOM:30.52 pp.
ICES. 2015. Report of the International Bottom Trawl Survey Working Group (IBTSWG), 23-27 March 2015, Bergen, Norway. ICES CM 2015/ SSGIEOM:20. 124 pp.
ICES. 2016a. Report of the Working Group on NephropsSurveys (WGNEPS), 7-8 November 2016, Reykjavík, Iceland. ICES CM 2016/ SSGIEOM:33.67 pp.
ICES. 2016b. Report of the Workshop on Nephrops Burrow Counting (WKNEP), 9-11 November 2016, Reykjavík, Iceland. ICES CM 2016/ SSGIEOM:33.65 pp.
ICES. 2016XX. EU request to ICES to provide Fmsy ranges for selected stocks in ICES subareas 5 to 10 . In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.2.3.1.
ICES. 2016YY. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.
ICES. 2016YY. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.
ICES. 2017. Report of the Working Group on Nephrops Surveys (WGNEPS). 28 November-1 December 2017, Heraklion, Greece. ICES CM 2017/SSGIEOM: 30. Ref: ACOM/ SCICOM.
ICES. 2018a. Report of the Working Group on Nephrops Surveys (WGNEPS), 6-8 November 2018. Lorient, France. ICES CM 2018/ EOSG:18.227 pp.
ICES. 2018b. Report of the Workshop on Nephrops burrow counting, 2-5 October 2018, Aberdeen, UK. ICES CM 2018/ EOSG:25.47 pp.
ICES. In prep. Manual for the Nephrops Underwater TV Surveys. Series of ICES Survey Protocols (SISP).
Leocádio, A., Weetman, A., and Wieland, K. (Eds). 2018. Using UWTV surveys to assess and advise on Nephrops stocks. ICES Cooperative Research Report No. 340. 49 pp. http://doi.org/10.17895/ices.pub. 4370.

Marrs S.J., Atkinson R.J.A., Smith C.J., Hills J.M. 1996. Calibration of the towed underwater TV technique for use in stock assessment of Nephrops norvegicus. Reference no. 94/069 (Study Project in support of the Common Fisheries Policy XIV/1810/C1/94, call for proposals 94/C 144/04).

Stokes, D., Gerritsen. H., O'Hea, B., Moore, S.J. and Dransfeld, L. 2014. "Irish Groundfish Survey Cruise Report, 24 September-17 December 2014", FEAS Survey Series; 2014/01. http://hdl.handle.net/10793/1064.

Table 22.2.1. Nephrops in FU19 (SW and SE Ireland). Landings in tonnes by country.

| Year | FU 19 |  |  |
| :--- | :---: | :---: | :---: |
|  | Rep. of Ireland |  |  |

Table 22.2.2. Nephrops in FU19 (SW and SE Ireland). Irish Nephrops directed effort (Kw Days) and landings.

| Year | Irish Fleet - Nephrops trawlers (>30\% landings weight) |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | All Vessels |  | Vessels >18 m |  |
|  | kW days ('000) | Landings Tonnes | kW days ('000) | Landings Tonnes |
| 1995 | 222.0 | 380 | 80.7 | 121 |
| 1996 | 178.6 | 355 | 55.6 | 86 |
| 1997 | 161.0 | 306 | 53.9 | 101 |
| 1998 | 329.6 | 498 | 144.6 | 189 |
| 1999 | 182.9 | 236 | 42.3 | 47 |
| 2000 | 142.0 | 217 | 56.2 | 86 |
| 2001 | 193.3 | 397 | 89.1 | 139 |
| 2002 | 506.7 | 883 | 323.7 | 446 |
| 2003 | 555.9 | 693 | 318.8 | 364 |
| 2004 | 488.1 | 558 | 303.0 | 311 |
| 2005 | 405.0 | 471 | 220.6 | 219 |
| 2006 | 424.2 | 478 | 208.8 | 186 |
| 2007 | 558.8 | 713 | 287.4 | 262 |
| 2008 | 534.1 | 643 | 288.1 | 319 |
| 2009 | 472.0 | 613 | 224.5 | 243 |
| 2010 | 382.2 | 494 | 103.7 | 114 |
| 2011 | 337.3 | 449 | 142.9 | 167 |
| 2012 | 355.5 | 541 | 91.9 | 126 |
| 2013 | 336.1 | 571 | 88.6 | 133 |
| 2014 | 213.6 | 332 | 52.1 | 74 |
| 2015 | 244.6 | 393 | 85.5 | 118 |
| 2016 | 287.3 | 558 | 111.2 | 233 |
| 2017 | 118.2 | 425 | 111.4 | 179 |
| 2018 | 71.6 | 107.1 | 24.1 | 29.9 |

Table 22.2.3. Nephrops in FU19 (SW and SE Ireland). Irish Sampling levels.

| Year | Quarter | Number of samples |  |  | Numbers Measured |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Landings | Catch | Discards | Landings |
| 2008 | 1 | 3 | 0 | 0 | 1502 | 0 | 0 |
| 2008 | 2 | 6 | 0 | 0 | 3521 | 0 | 0 |
| 2008 | 3 | 6 | 0 | 0 | 6412 | 0 | 0 |
| 2008 | 4 | 3 | 0 | 0 | 876 | 0 | 0 |
| 2009 | 1 | 3 | 0 | 0 | 1347 | 0 | 0 |
| 2009 | 2 | 6 | 0 | 0 | 3369 | 0 | 0 |
| 2009 | 3 | 2 | 0 | 0 | 1003 | 0 | 0 |
| 2009 | 4 | 5 | 0 | 0 | 1882 | 0 | 0 |
| 2010 | 1 | 2 | 0 | 0 | 840 | 0 | 0 |
| 2010 | 2 | 7 | 0 | 0 | 2989 | 0 | 0 |
| 2010 | 3 | 4 | 0 | 0 | 1457 | 0 | 0 |
| 2010 | 4 | 6 | 0 | 0 | 2376 | 0 | 0 |
| 2011 | 1 | 3 | 0 | 0 | 1493 | 0 | 0 |
| 2011 | 2 | 5 | 0 | 0 | 2747 | 0 | 0 |
| 2011 | 3 | 2 | 0 | 0 | 938 | 0 | 0 |
| 2011 | 4 | 5 | 0 | 0 | 2686 | 0 | 0 |
| 2012 | 1 | 6 | 0 | 0 | 2053 | 0 | 0 |
| 2012 | 2 | 7 | 0 | 0 | 3956 | 0 | 0 |
| 2012 | 3 | 4 | 0 | 0 | 1980 | 0 | 0 |
| 2012 | 4 | 4 | 0 | 0 | 1969 | 0 | 0 |
| 2013 | 1 | 3 | 0 | 0 | 1857 | 0 | 0 |
| 2013 | 2 | 8 | 5 | 0 | 4117 | 2059 | 0 |
| 2013 | 2 | 3 | 3 | 0 | 1177 | 1250 | 0 |
| 2013 | 4 | 3 | 3 | 0 | 1472 | 1276 | 0 |
| 2014 | 1 | 3 | 2 | 0 | 1137 | 941 | 0 |
| 2014 | 2 | 7 | 7 | 0 | 3331 | 2319 | 0 |
| 2014 | 3 | 3 | 2 | 0 | 1344 | 682 | 0 |
| 2014 | 4 | 10 | 8 | 0 | 3455 | 2200 | 0 |
| 2015 | 1 | 1 | 1 | 0 | 417 | 310 | 0 |
| 2015 | 2 | 3 | 3 | 0 | 1417 | 1267 | 0 |
| 2015 | 3 | 2 | 2 | 1 | 856 | 648 | 321 |
| 2015 | 4 | 3 | 2 | 0 | 1250 | 774 | 0 |
| 2016 | 1 | 3 | 3 | 0 | 1500 | 1631 | 0 |
| 2016 | 2 | 6 | 5 | 0 | 2310 | 1760 | 0 |
| 2016 | 3 | 9 | 7 | 0 | 3328 | 2448 | 0 |
| 2016 | 4 | 5 | 5 | 0 | 1,923 | 1521 | 0 |

Table 22.2.3. Continued.

| Year | Quarter | Number of samples |  | Numbers Measured |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2017 | 1 | Catch | Discards | Landings | Catch | Discards | Landings |
| 2017 | 2 | 4 | 4 | 0 | 1860 | 1283 | 0 |
| 2017 | 3 | 2 | 3 | 0 | 1572 | 1281 | 0 |
| 2017 | 4 | 4 | 2 | 0 | 998 | 943 | 0 |
| 2018 | 1 | 7 | 7 | 0 | 300 | 785 | 0 |
| 2018 | 2 | 1 | 1 | 0 | 3579 | 3230 | 0 |
| 2018 | 3 | 4 | 1 | 0 | 255 | 275 | 0 |
| 2018 | 4 | 2 | 0 | 370 | 404 | 0 |  |

Table 22.2.4. Nephrops in FU19 (SW and SE Ireland). Landings and estimated discards by weight and numbers.

| Year | Female '000s <br> Landings ( t ) | Male '000s |  |  | Both sexes <br> \% Discard |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Discards (t) | Landings ( t ) | Discards (t) |  |
| 2008 | 99 | 29 | 691 | 69 | 11 |
| 2009 | 117 | 106 | 681 | 141 | 24 |
| 2010 | 138 | 98 | 522 | 148 | 27 |
| 2011 | 169 | 155 | 450 | 250 | 39 |
| 2012 | 190 | 202 | 647 | 265 | 36 |
| 2013 | 259 | 210 | 525 | 220 | 35 |
| 2014 | 106 | 71 | 353 | 87 | 26 |
| 2015 | 79 | 64 | 423 | 101 | 25 |
| 2016 | 154 | 91 | 429 | 100 | 25 |
| 2017 | 133 | 58 | 280 | 79 | 25 |
| 2018 | 64 | 25 | 146 | 38 | 23 |
| 2008 | 3,893 | 1,781 | 19,516 | 3,255 | 18 |
| 2009 | 5,819 | 8,250 | 20,324 | 8,793 | 39 |
| 2010 | 6,276 | 8,147 | 16,001 | 10,117 | 45 |
| 2011 | 7,295 | 12,895 | 16,900 | 18,192 | 56 |
| 2012 | 9,266 | 17,635 | 22,540 | 19,108 | 54 |
| 2013 | 11,680 | 18,945 | 17,399 | 17,034 | 55 |
| 2014 | 4,862 | 5,647 | 11,183 | 5,572 | 41 |
| 2015 | 3,706 | 5,255 | 13,111 | 6,462 | 41 |
| 2016 | 6,877 | 6,761 | 12,610 | 6,668 | 41 |
| 2017 | 5,295 | 4,400 | 9,022 | 5,044 | 40 |
| 2018 | 2,617 | 1,692 | 4,818 | 2,279 | 35 |

Table 22.2.5. Nephrops in FU19 (SW and SE Ireland). Area ( $\mathrm{Km}^{2}$ ) of discrete patches and percentage contribution to overall area.

| Ground | Area ( $\left.\mathrm{Km}^{2}\right)$ | \% Contribution |
| :--- | :---: | :---: |
| Bantry | 121.5 | $6 \%$ |
| Cork Channels | 562.0 | $28 \%$ |
| Galley Grounds 1 | 60.9 | $3 \%$ |
| Galley Grounds 2 | 76.7 | $4 \%$ |
| Galley Grounds 3 | 133.9 | $47 \%$ |
| Galley Grounds 4 | 925.1 | $2 \%$ |
| Helvick 1 | 33.1 | $3 \%$ |
| Helvick 2 | 59.5 | $4 \%$ |
| Total | 1972.8 | $4 \%$ |

Table 22.2.6. Nephrops in FU19 (SW and SE Ireland). Detailed summary statistics for the various Nephrops patches in FU19 over the time-series. ( $\mathrm{N}=$ number of stations, Mean Density ( $\mathrm{no} / \mathrm{m}^{2}$ ) is adjusted for the bias correction factor in Table 3, sd, se and ci are the standard deviation, standard error and $95 \%$ confidence intervals on the mean density).

| Year | Ground | N | Mean Density ( $\mathrm{no} / \mathrm{m}^{2}$ ) | sd | se | ci |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | Gallev Grounds 4 | 6 | 0.21 | 0.18 | 0.08 | 0.19 |
| 2011 | Bantry | 5 | 0.33 | 0.23 | 0.1 | 0.28 |
| 2011 | Cork Channels | 12 | 0.35 | 0.32 | 0.09 | 0.2 |
| 2011 | Gallev Grounds 1 | 3 | 0.52 | 0.41 | 0.24 | 1.02 |
| 2011 | Galley Grounds 2 | 3 | 0.59 | 0.43 | 0.25 | 1.07 |
| 2011 | Gallev Grounds 3 | 4 | 0.58 | 0.22 | 0.11 | 0.35 |
| 2011 | Helvick 1 | 3 | 0.6 | 0.01 | 0.01 | 0.04 |
| 2011 | Helvick 2 | 5 | 0.12 | 0.21 | 0.09 | 0.26 |
| 2012 | Bantrv | 1 | 0.2 | NA | NA | NA |
| 2012 | Cork Channels | 9 | 0.27 | 0.17 | 0.06 | 0.13 |
| 2012 | Galley Grounds 2 | 4 | 0.59 | 0.12 | 0.06 | 0.19 |
| 2012 | Gallev Grounds 3 | 1 | 0.51 | NA | NA | NA |
| 2012 | Gallev Grounds 4 | 16 | 0.39 | 0.16 | 0.04 | 0.09 |
| 2012 | Helvick 1 | 3 | 0.33 | 0.13 | 0.08 | 0.33 |
| 2012 | Helvick 2 | 6 | 0.33 | 0.41 | 0.17 | 0.43 |
| 2013 | Bantrv | 4 | 0.38 | 0.2 | 0.1 | 0.31 |
| 2013 | Cork Channels | 11 | 0.12 | 0.1 | 0.03 | 0.07 |
| 2013 | Gallev Grounds 1 | 2 | 0.23 | 0.18 | 0.13 | 1.59 |
| 2013 | Gallev Grounds 2 | 3 | 0.48 | 0.44 | 0.25 | 1.09 |
| 2013 | Gallev Grounds 3 | 4 | 0.59 | 0.24 | 0.12 | 0.38 |
| 2013 | Gallev Grounds 4 | 13 | 0.19 | 0.27 | 0.07 | 0.16 |
| 2013 | Helvick 1 | 1 | 0.09 | NA | NA | NA |
| 2013 | Helvick 2 | 2 | 0.06 | 0.05 | 0.04 | 0.48 |
| 2014 | Bantrv | 4 | 0.25 | 0.05 | 0.03 | 0.09 |
| 2014 | Cork Channels | 10 | 0.1 | 0.06 | 0.02 | 0.04 |
| 2014 | Galley Grounds 1 | 2 | 0.61 | 0.41 | 0.29 | 3.69 |
| 2014 | Gallev Grounds 2 | 2 | 0.82 | 0.14 | 0.1 | 1.23 |
| 2014 | Gallev Grounds 3 | 4 | 0.66 | 0.23 | 0.12 | 0.37 |
| 2014 | Galley Grounds 4 | 14 | 0.29 | 0.29 | 0.08 | 0.17 |
| 2014 | Helvick 1 | 2 | 0.67 | 0.28 | 0.2 | 2.53 |
| 2014 | Helvick 2 | 2 | 0.03 | 0.04 | 0.03 | 0.39 |
| 2015 | Bantry | 2 | 0.32 | 0.11 | 0.08 | 1.02 |
| 2015 | Cork Channels | 10 | 0.08 | 0.11 | 0.03 | 0.08 |
| 2015 | Gallev Grounds 1 | 2 | 0.32 | 0.46 | 0.32 | 4.12 |
| 2015 | Galley Grounds 2 | 2 | 0.53 | 0.08 | 0.06 | 0.74 |
| 2015 | Gallev Grounds 3 | 4 | 0.40 | 0.14 | 0.07 | 0.23 |
| 2015 | Gallev Grounds 4 | 14 | 0.27 | 0.19 | 0.05 | 0.11 |
| 2015 | Helvick 1 | 2 | 0.30 | 0.23 | 0.16 | 2.08 |
| 2015 | Helvick 2 | 2 | 0.09 | 0.09 | 0.06 | 0.79 |
| 2015 | Kenmare Bav | 1 | 0.30 | NA | NA | NA |

Table 22.2.6. Continued

| Year | Ground | $N$ | Mean Densitv ( $\mathrm{no} / \mathrm{m}^{2}$ ) | sd | se | ci |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | Bantrv | 4 | 0.20 | 0.07 | 0.04 | 0.12 |
| 2016 | Cork Channels | 10 | 0.21 | 0.11 | 0.03 | 0.08 |
| 2016 | Galley Grounds 1 | 2 | 0.03 | 0.01 | 0.01 | 0.08 |
| 2016 | Gallev Grounds 2 | 2 | 0.53 | 0.12 | 0.09 | 1.11 |
| 2016 | Gallev Grounds 3 | 4 | 0.16 | 0.12 | 0.06 | 0.19 |
| 2016 | Galley Grounds 4 | 14 | 0.17 | 0.20 | 0.05 | 0.12 |
| 2016 | Helvick 1 | 2 | 0.38 | 0.08 | 0.06 | 0.70 |
| 2016 | Helvick 2 | 2 | 0.07 | 0.09 | 0.06 | 0.81 |
| 2016 | Kenmare Bay | 2 | 0.24 | 0.15 | 0.11 | 1.33 |
| 2017 | Bantrv | 3 | 0.29 | 0.15 | 0.09 | 0.37 |
| 2017 | Cork Channels | 10 | 0.25 | 0.20 | 0.06 | 0.14 |
| 2017 | Galley Grounds 1 | 2 | 0.24 | 0.11 | 0.08 | 1.00 |
| 2017 | Gallev Grounds 2 | 2 | 0.63 | 0.06 | 0.04 | 0.55 |
| 2017 | Gallev Grounds 3 | 3 | 0.45 | 0.12 | 0.07 | 0.30 |
| 2017 | Gallev Grounds 4 | 15 | 0.16 | 0.16 | 0.04 | 0.09 |
| 2017 | Helvick 1 | 2 | 0.46 | 0.07 | 0.05 | 0.66 |
| 2017 | Helvick 2 | 2 | 0.16 | 0.23 | 0.16 | 2.03 |
| 2017 | Kenmare Bav | 2 | 0.16 | 0.22 | 0.16 | 1.97 |
| 2018 | Bantrv | 4 | 0.06 | 0.02 | 0.01 | 0.04 |
| 2018 | Cork Channels | 10 | 0.11 | 0.11 | 0.04 | 0.08 |
| 2018 | Gallev Grounds 1 | 2 | 0.06 | 0.01 | 0.01 | 0.10 |
| 2018 | Gallev Grounds 2 | 2 | 0.19 | 0.19 | 0.14 | 1.75 |
| 2018 | Gallev Grounds 3 | 4 | 0.11 | 0.09 | 0.05 | 0.14 |
| 2018 | Gallev Grounds 4 | 14 | 0.07 | 0.08 | 0.02 | 0.05 |
| 2018 | Helvick 1 | 2 | 0.11 | 0.10 | 0.07 | 0.92 |
| 2018 | Helvick 2 | 2 | 0.06 | 0.03 | 0.02 | 0.28 |
| 2018 | Kenmare Bay | 2 | 0.07 | 0.03 | 0.02 | 0.25 |
| 2019 | Bantrv | 4 | 0.13 | 0.04 | 0.02 | 0.06 |
| 2019 | Cork Channels | 10 | 0.16 | 0.17 | 0.06 | 0.13 |
| 2019 | Galley Grounds 1 | 2 | 0.12 | 0.17 | 0.12 | 1.57 |
| 2019 | Gallev Grounds 2 | 2 | 0.66 | 0.38 | 0.27 | 3.40 |
| 2019 | Gallev Grounds 3 | 4 | 0.21 | 0.14 | 0.07 | 0.23 |
| 2019 | Galley Grounds 4 | 14 | 0.18 | 0.23 | 0.06 | 0.13 |
| 2019 | Helvick 1 | 2 | 0.34 | 0.27 | 0.19 | 2.46 |
| 2019 | Helvick 2 | 2 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2019 | Kenmare Bay | 2 | 0.27 | 0.10 | 0.07 | 0.88 |
| 2019 | Dunmanus Bav | 2 | 0 | 0 | 0 | 0 |

Table 22.2.7. Nephrops in FU19 (SW and SE Ireland). Summary statistics for FU19 combined over the timeseries. No TV survey from 2007-2010.

| Year | Number of <br> stations | Mean Density <br> adjusted (bur- <br> row $\mathbf{m}^{2}$ ) | Standard <br> Deviation | Raised abundance <br> estimate adjusted <br> (million burrows) | Upper 95\%CI <br> on Abundance | Lower 95\%CI <br> on Abundance | CVs |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 22.3.1. Nephrops in FU19 (SW and SE Ireland). Forecast inputs (bold) and historical estimates of mean weight in landings and harvest rate (landings + dead discards)/(abundance estimate), discard rate (discards divided by landings + discards) and dead discard rate as dead discards divided by removals (landings + dead discards).

| Year | Landings in number | Total discards* in number | Removals in number | Discard Rate number | Dead discard rate number | UWTV abundance estimate | 95\% Conf. intervals | Harvest rate | Landings | Total discards* | Mean weight in landings | Mean weight in discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions | millions | millions | \% | \% | millions | millions | \% | tonnes | tonnes | grammes | grammes |
| 2006 | 26.2 | 2.6 | 28.1 | 8.9 | 6.8 | na | na | na | 741 | 37 | 28.3 | 14.4 |
| 2007 | 30.8 | 1.5 | 31.9 | 4.8 | 3.6 | na | na | na | 957 | 26 | 31.1 | 17 |
| 2008 | 25.2 | 5.4 | 29.3 | 17.7 | 13.9 | na | na | na | 866 | 107 | 33.7 | 19.3 |
| 2009 | 28.4 | 18.5 | 42.3 | 39.5 | 32.8 | na | na | na | 833 | 258 | 30.5 | 14.5 |
| 2010 | 23.2 | 19.0 | 37.4 | 45.1 | 38.1 | na | na | na | 722 | 269 | 29.6 | 13.5 |
| 2011 | 25.8 | 32.4 | 50.1 | 55.7 | 48.5 | 665 | 171 | 7.10 | 608 | 387 | 25 | 12.6 |
| 2012 | 32.3 | 37.3 | 60.2 | 53.6 | 46.4 | 594 | 111 | 9.10 | 770 | 420 | 26.4 | 12.7 |
| 2013 | 29.5 | 36.5 | 56.8 | 55.3 | 48.1 | 487 | 161 | 11.00 | 781 | 404 | 27.4 | 12.1 |
| 2014 | 16.3 | 11.4 | 24.9 | 41.1 | 34.4 | 636 | 188 | 3.90 | 468 | 161 | 28.6 | 14.1 |
| 2015 | 17.0 | 11.8 | 25.9 | 41.1 | 34.3 | 482 | 126 | 5.50 | 507 | 177 | 29.8 | 13.8 |
| 2016 | 19.7 | 13.6 | 29.9 | 40.8 | 34.1 | 399 | 100 | 7.50 | 591 | 194 | 29.9 | 14.2 |
| 2017 | 14.6 | 9.6 | 21.8 | 39.7 | 33.1 | 499 | 120 | 4.4 | 420 | 138 | 28.8 | 14.45 |
| 2018 | 7.8 | 4.2 | 10.9 | 34.8 | 28.6 | 176 | 53 | 6.2 | 219 | 65 | 28.2 | 15.7 |
| 2019 |  |  |  |  |  | 386 | 127 |  |  |  |  |  |
|  |  |  | Average 16-18 | 38.5 | 31.9 |  |  |  |  | Average 16-18 | 29.0 | 14.8 |



Figure 22.1.1. Nephrops in FU19 (Ireland SW and SE Coast). Revised discrete patches overlaid on overlaid on proportion of Nephrops in the Irish landings overlaid on international OTB effort (red=0\% Nephrops; blue=5060\% Nephrops; grey=unknown (no Irish landings).


Figure 22.1.2. Nephrops in FU19 (Ireland SW and SE Coast). Time-series of the number of Irish vessels reporting landings of Nephrops from FU19 with a $>10 \mathrm{t}$ threshold.


Figure 22.1.3. Nephrops in FU19 (Ireland SW and SE Coast). Combined box and kite plot of vessel power by year. The blue line indicates the mean.


Figure 22.2.1. Nephrops in FU19 (Ireland SW and SE Coast). Landings in tonnes by country.


Figure 22.2.2. Nephrops in FU19 (Ireland SW and SE Coast). Trawl effort for Irish OTB vessels where $>30 \%$ of landed weight was Nephrops.


Figure 22.2.3. Nephrops in FU19 (Ireland SW and SE Coast). Sampling levels for FU19.


Figure 22.2.4. Nephrops in FU19 (Ireland SW and SE Coast). Mean size trends for catches (dotted) and whole landings (solid) by sex 2002-2018. Vertical lines displayed are Minimum Conservation Reference Size 25 mm Carapace Length (CL) and 35 mm CL.


Figure 22.2.5. Nephrops in FU19 (Ireland SW and SE Coast). Annual sex ratio of landings (2008-2018) and catch (2008-2018).


Figure 22.2.6. Nephrops in FU19 (Ireland SW and SE Coast). Mean weight in Bantry Bay catch samples by sex with loess smoother and showing cyclical trends.


Figure 22.2.7. Nephrops in FU19 (Ireland SW and SE Coast). Mean weight in catch data for all grounds in FU19 by sex with loess smoother and showing cyclical trends.


Figure 22.2.8. Nephrops in FU19 (Ireland SW and SE Coast). Annual estimated mean weights (gr) in the landings and discards.


Figure 22.2.9. Nephrops in FU19 (Ireland SW and SE Coast). Violin and box plot a of adjusted burrow density (burrow $/ \mathrm{m}^{2}$ ) distributions by year from 2006-2019. The blue line indicates the mean density over time. The horizontal black line represents the median, white box is the interquartile range, the black vertical line is the range and the black dots are outliers. No estimate available for Galley Ground 4 in 2011, Galley Ground 1 in 2012. No TV survey from 2007 to 2010.


Figure 22.2.10. Nephrops in FU19 (Ireland SW and SE Coast). Time-series of total abundance estimates for FU19 (error bars indicate $95 \%$ confidence intervals) and $B_{\text {trigger }}$ is dashed green line.

## Length frequencies for IGFS Survey Catches: <br> Nephrops in FU19



Figure 22.2.11. Nephrops in FU19 (Ireland SW and SE Coast). Mean size trends for catches by sex from Irish Groundfish Survey 2003-2018. Vertical lines displayed are Minimum Conservation Reference Size 25 mm Carapace Length (CL) and 35 mm CL.


Figure 22.3.1. Nephrops in FU19 (Ireland SW and SE Coast). Harvest Rate (\% dead removed/UWTV abundance). The dashed and solid lines are the MSY proxy and the harvest rate respectively.

# 20 Norway lobster (Nephrops norvegicus) in divisions 7.g and 7.h, Functional Units 20 and 21 (Celtic Sea) 

## Type of assessment in 2019

A full UWTV based assessment was carried out and catch options based on the new stock-specific reference points estimated by WGCSE 2016 using the methods applied to other Nephrops stocks at WKFMSYREF4 (ICES, 2016).

## ICES advice applicable to 2018

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2014-2016, catches in 2018 should be no more than 8673 tonnes.

To ensure that the stock in functional unit (FU) 20 and 21 is exploited sustainably, management should be implemented at the functional unit level."

## ICES advice applicable to 2019

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2015-2017, catches in 2019 should be no more than 5320 tonnes.

To ensure that the stock in functional units 20 and 21 is exploited sustainably, management should be implemented at the level of the combined functional units 20 and 21."

### 20.1 General

## Stock description and management units

The FU20-21 Nephrops stock is included in the whole ICES Area 7 together with Irish Sea East and West [FU14, FU15], Porcupine Bank [FU16], Aran Islands [FU17], northwest Irish Coast [FU18], southeast and southwest Irish Coast [FU19], Smalls [FU22]. The TAC is set for Subarea 7 , which does not correspond to the stock area.

Historically FU20-22 fishery and sampling data covered an amalgamation of several spatially distinct mud patches; FU20 NW Labadie, Baltimore and Galley, FU21 Jones and Cockburn and FU22 the Smalls. WGCSE 2013 recommended that FU20-22 should be split into FU20-21 combined and FU22 for the purposes of assessment and advice provision. There is evidence that the Celtic Sea Nephrops patches are linked in meta-population sense (O'Sullivanet al., 2015). However, fishing mortality and biological parameters (density, growth, $M$, etc.) may vary across the different patches. The map below shows FU20-21 assessment area (blue) and TAC area (red). There is no evidence that the individual functional units belong to the same stock. See Section 18 for details on Nephrops in Subarea 7 general section.


## Ecosystem aspects

Details of the ecosystem on FU20-21 are provided in the stock annex updated by WKCELT.

## Fishery description

Ireland, France and the UK are the main countries involved in the FU20-21 Nephrops fishery. In the early 2000s, the Republic of Ireland fleet had on average $10 \%$ of the landings and this has increased to approximately $60 \%$ from this FU in recent times. A description of this fleet is given in the stock annex. The fishery on FU20-21 grounds operates throughout the year, weather permitting with a seasonal trend and has expanded in the mid-2000s. In 2011, Irish landings were higher than French landings for the first time. The time-series of numbers of vessels with landings greater than 10 tonnes is updated in Figure 23.1.1. The time-series of vessel power is shown as a box and kite plot in Figure 23.1.2. In recent years, the Irish fleet have increased landings from the southern part of the grounds (see stock annex).

French trawlers targeting Nephrops in the Celtic Sea operate mainly in the FU20-21 component of the stock. France dominated in the landings in the early 2000s on average $90 \%$ of landings and this has decreased to about $20 \%$ in recent times. A description of this fleet is given in the stock annex.

There is an increase in participation by the UK in this fishery in the most recent years The UK fleet had on average $20 \%$ of the landings from this FU in recent times (2016-2018) with highest landings recorded in 2016 ( 445 t ).

## Fishery in 2018

## Ireland

In recent years, several newer vessels specializing in Nephrops fishing have participated periodically in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates. There has been a trend for Irish vessels to switch to multi (quad)rig trawls since 2012. Thesevessels are more efficient atcatching Nephrops (BIM, 2015).

In 2018, 53 vessels reported landings in excess of 10 t accounting for $92 \%$ of total Irish landings.

## France

In 2018, 36 French vessels reported landings from FU20-21 where many of these switch between FU20-21 and FU22 within a trip.

## UK

18 UK(E\&W) vessels reported landings for FU20-21, seven vessels from Scotland and four vessels from Northern Ireland.

## Information from stakeholders

None presented.

### 20.2 Data

## InterCatch

Data were available in InterCatch and used, and used for catch data only. French data provided directly by the national expert and not extracted from InterCatch.

## Landings

The reported landings time-series is shown in Figure 23.2.1 and Table 23.2.1.
The reported Irish landings from FU20-21 have increased since the mid-2000s to the second highest in the Irish time-series in 2016 (1531 t) although have decreased in 2018 to 1197 t. French landings have gradually decreased since the early 2000s to the present lowest value of 195 t . Reported landings from the UK have fluctuated with an increasing trend since 2015. UKEngland \& Wales had the highest landings at 348 t followed by Scotland ( 34 t ), Northern Ireland reporting 28 t , and minor landings from Belgium less than 0.2 t .
The overall fishing profile remains typically seasonal with the majority of the Irish and UK landings coming from the second quarter (see stock annex).

## Effort

Effort data are available for the Irish Nephrops directed fleet in FU2021 from 1995-2018. The effort series is based on the same criteria for FU15, 16, 17, 19and 22 ( $30 \%$ landings threshold) and will be contingent on the accuracy of landings data reported in logbooks. Effort data are not standardized, and hence do not take into account vessel capabilities, efficiency, seasonality or other factors that maybias perception of LPUE as an abundance trend over the longer term. These data are not used in the assessment.

WGCSE 2015 recommended that effort data in Kw days should be presented as these data are more informative that uncorrected effort data. Effort data are available from 1995 for the Irish otter trawl Nephrops directed fleet. In 2018, this fleet accounted for $\sim 90 \%$ of the landings compared with an average of $70 \%$ over the time period. Effort shows an increasing trend since the mid-2000s to a sharp decrease since 2015 (Figure 23.2.2 and Table 23.2.2).

Effort data in KW days are not available for France. Previously effort data were reported from 1983 to 2008 for the French Nephrops fleet for the combined Celtic Sea FU20-22 (see stock annex). Since 2009, a new registration system of official French statistics has changed the way fishing effort is computed, and a new threshold method of 500 kg landed by trip is used to report effort. French fishing effort reported in hours and LPUE ( $\mathrm{kg} / \mathrm{hr}$ ) since 2009 shows an overall declining trend (Table 23.2.3).

## Sampling levels

Sampling levels, data aggregating and raising procedures were reviewed by WKCELT 2014 and are documented in the stock annex. The time-series of samples is shown in Table 23.2.4, and remains sparse due to the offshore nature of the fishery although progress is being made by Ireland.

## Commercial length-frequency distributions

Prior to 2012 there was insufficient Irish sampling to generate length-frequency distributions although since then efforts are being progressed. For France limited data were available for 1997 and 2010-2013 (see stock annex for details).

Length-frequency distributions of landings and discards for both countries from 2012 to 2018 are presented in Figure 23.2.3 along with the European minimum conservation referencesize ( 25 CL mm) and French ( 35 CL mm ) minimum landings size also shown.

The short series on LFDs for both countries shows that the LFDs differ between the two countries. The French fishery caught higher proportions of larger individuals ( $>35 \mathrm{~mm}$ ) on average $70 \%$ compared to $41 \%$ for the Irish fishery for the available time-series.

## Sex ratio

The sex ratio is male biased from the available French and Irish sampling data (Table 23.2.5).

## Mean weight explorations

The numbers in the French landings and discards raised to FU20-21 only for 2012-2016 were provided to WGCSE 2017. These data (years 2012-2015) are similar to that reported by WGCSE 2015, which could not be reproduced at WGCSE 2016. At WGCSE 2016, a scaling factor was applied to the French dataset as these were provided raised to the whole of area FU20-22. The French dataset provided toWGCSE2017 (years2012-2015) results in an increase in mean weights and decrease in removals from that previously reported at WGCSE 2016 (Table 23.2.6). The working group accepted the French dataset, and this is used to calculate the estimated annual mean weights in the landings and discards.

WGCSE2019 used the length-weight relationship as described in stock annexto raise both countries sampling data, which are based on Scottish data (Pope and Thomas, 1955).

The mean weight in the landings for France is higher than that in the Irish landings (Table 23.2.7). The estimated annual mean weights in the landings and discards by country and also combined scaled to the international landings is shown in Table 23.2.8 and Figure 23.2.4).

## Discards

For the Irish data, discard rates have been estimated using unsorted catchand discards sampling. This involves unsorted catch and discard samples being provided by vessels or collected by observers at-sea on discard trips. The catch sample is partitioned into landings and discards using an on-board discard selection ogive derived for the discard samples. Due to sparse sampling effort annual aggregations are used to derive length distributions and selection ogives. Figure 23.2.5 shows the annual discard ogive from the Irish sampling used to partition the catch. The length-weight regression parameters given in the stock annex are used to calculate sampled weights and appropriate annual raising factors. The sampling intensity and coverage has varied over the short time-series, and is relatively poor compared to other Nephropsstocks but at present it is the best available.

Estimated discard rates range between $18-41 \%$ of total catch by number and 10-27\% of total catch by weight in the Irish fishery shown in Table 23.2.7. In the French fishery, estimated discard rates range between $25-78 \%$ of total catch by number and $16-56 \%$ of total catch by weight shown in Table 23.2.6.

Estimated discard rates for both countries combined in shown in Table 23.2.8 and these range between $24-52 \%$ of total catch by number and $14-31 \%$ of total catch by weight. Discard rate of females tends to be higher due to the smaller average size and market reasons as is observed in other Nephrops fisheries.

There is no information on discard survival rate in this fishery. $25 \%$ is assumed in line with other Nephrops stocks in the Celtic Sea (Charuau et al., 1982).

Gear selectivity trials by Bord Iascaigh Mhara (BIM, 2017) reported a $64 \%$ survivor rate for Nephrops caught in a trawl with a SELTRA selectivity device in the outer Galway Bay area.

Table 23.3.1 gives weights, numbers and mean weights of the landings and discard raised internationally according to the stock annex.

## Abundance indices from UWTV surveys

The methods used during the survey were similar to those employed for UWTV surveys of Nephrops stocks around Ireland and elsewhere and are documented by WKNEPHTV (ICES, 2007), SGNEPS (ICES, 2009; 2010; 2012), WGNEPS (ICES, 2013; 2014; 2015; 2016a; 2017; 2018a), WKNEPS (ICES, 2016b;2018b) and Leocádio et al., 2018. SGNEPS (ICES, 2012) recommended that a CV (or relative standard error) of $<20 \%$ is an acceptable precision level for UWTV survey estimates of abundance. UWTV surveys conducted in 2006 and 2012 are deemed exploratory as stations were chosen based on areas heavily fished by vessels (Doyleet al., 2013). These are likely to be biased estimate of density and cannot be extrapolated to estimate density for the whole area. A randomised isometric grid design was employed with UWTV stations at 6.0 nmi intervals for 2013-2018 surveys. The 2013 survey achieved partial coverage $\sim 60 \%$ of the total area. The 2013 abundance has been scaled up to the entire area since densities in the un-surveyed part of the ground were not significantly different in 2014. From 2014 to 2018, full survey coverage was achieved. The geo-statistical analysis for years 2013 to 2019 follows the steps documented in White et al., 2019.

The 2019 mean burrow density was 0.06 burrows $/ \mathrm{m}^{2}$ compared with 0.27 burrows $/ \mathrm{m}^{2}$ in 2018 . The 2019 geostatistical abundance estimate was 617 million a $77 \%$ decrease on the abundance for 2018 with a CV of $5 \%$, which is well below the upper limit of $20 \%$ recommended by SGNEPS 2012. There was a general decrease in densities observed in 2019. Figure 23.2 .6 shows the krigged contour and density plots for the time-series. The summary statistics from this geostatistical analysis are given in Table 23.2.9 and plotted in Figure 23.2.7. The estimation variance of the survey is very low (CVs in the order 5\%).
The Nephrops Underwater TV Surveys SISP guidelines (ICES, in prep) were followed in 2019 given the substantial decrease observed. A random selection of $20 \%$ of UWTV stations were reviewed in thelaboratory following the same procedures that are carried out on board. Full details are available in R-markdown (ICES, 2019 Annex 3, WD 10). The results showed an overall increase ( $15.5 \%$ ) in the review counts for these selected stations comparing them with the on-board counts. The review process also confirmed the observed low density estimates in 2019. The onboard count data are used to calculate the abundance estimate for determining catch scenarios for 2020.

## Groundfish survey data

There are two IBTS-GFS catching Nephrops in FU20-21: French groundfish survey EVHOE-WI-BTS-Q4 since 1997 and Irish groundfish survey-Q4: IGFS-WIBTS-Q4 commenced in 2003 (Stokes et al., 2014). These provide information on length-frequency compositions, mean size in the catches, CPUE of Nephrops in FU20-21 (ICES, 2015). The mean size of the catches is stable over the time-series except in 2006 and 2008, which signals recruitment into the fishery in 2006 and 2007 as shown by the Irish IBTS survey in Figure 23.2.8 and the French IBTS survey (Figure 23.2.9). There is no 2017 length dataset for EVHOE due to research vessel breakdown.

### 20.3 Assessment

## Comparison with previous assessments

The WGCSE 2019 carried out a full UWTV based assessment for this stock using the stock stockspecific reference points were estimated by the 2016 working group based on methods for other Nephropsstocks used by WKMS YREF4 (ICES, 2016). This is in accordance with recommendations by WKCELT 2014 where data improvements have been made for this stock such as:

- complete survey coverage of the stock area giving quality assured density estimates and abundance estimates conforming to WGNEPS recommendations; and also
- improved sampling data achieving better coverage and robust estimates of the various parameters need to calculate catch options (e.g. mean weight in the landings and discards, discard percentage in numbers).


## State of the stock

UWTV abundanceestimates suggest that the stock size has fluctuated over the short time-series. The 2019 estimate is a decrease from 2018 estimate by $77 \%$.

No MSY $B_{\text {trigger }}$ has been proposed as the time-series is too short (six years of full TV survey coverage).

Table 23.3.1 and Figure 23.3 .1 summarize recent harvest ratios, which have been below the FmsY proxy for the last three years.

### 20.4 Catch scenario table

Catch scenario table inputs and estimates of mean weight in landings and harvest ratios are presented in Table 23.3.1 and summarised below.

In line with previous practice, an average(2016-2018) of mean weights is used to account for this variability. Three-year average (2016-2018) of proportion of removals retained was used as is standard for other Nephrops stocks.

The basis for the catch scenario.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Stock abundance | 617 mil- <br> lion indi- <br> viduals | UWTV survey 2019 |
| Mean weight in <br> wanted catch | 33.1 g | Average 2016-2018 |
| Mean weight in un- <br> wanted catch | 18.3 g | Average 2016-2018 |
| Unwanted catch rate | $29.2 \%$ | Average 2016-2018 (by number). Calculated as discards divided by landings + dis- <br> cards. |
| Discard survival rate | $25 \%$ | Only applies in scenarios where discarding is allowed. <br> Dead unwanted catch <br> rate |

A prediction of landings for the FU20-21 using the approach agreed procedure proposed at WKNEPH 2009 and outlined in the stock annex will be made on the basis of the 2019 UWTV survey. This will be presented in October 2019 for the provision of advice.

### 20.5 Reference points

New reference points were estimated by WGCSE 2016 using the same method and approach used at WKMS YREF4 (ICES, 2016). The detailed analysis is available in working document 11. In the case of FU20-21 there is a limited number of years for which length-frequency data were available, so the three-year moving window could only be applied to give two estimates. The resulting potential $\mathrm{F}_{\mathrm{MS}}$ harvest rates and ranges are given in the following table.

| YEAR | FMAX | FMAX.LOW | FMAX.UP | F35 | F35.LOW | F35.UP | F0.1 | FO.1.LOW | F0.1.UP |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 9.12 | 6.51 | 12.60 | 11.03 | 6.11 | 13.21 | 5.91 | 5.08 | 15.11 |
| 2013 | 9.45 | 6.71 | 13.26 | 11.17 | 6.30 | 13.78 | 6.10 | 5.23 | 15.93 |

Given the low density in the area and combined sex $\mathrm{F}_{0.1}$ was considered and appropriate $\mathrm{F}_{\mathrm{MS}}$ proxy.

| STOCK CODE | MSY FLOWER* | FMSY* | MSY FUPPER*WITH AR | MSY BTRIGGER | MSY FUPPER*WITH NO AR |
| :--- | :---: | :---: | :---: | :---: | :---: |
| nep-2021 | $5.9 \%$ | $6.0 \%$ | $6.0 \%$ | Not defined | $6.0 \%$ |

* Harvest rate (HR).

No proposal has been made for MSY Btrigger as the time-series is too short.

### 20.6 Management plans

There is no specific management plan for the FU20-21 Nephrops.
The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to Norway lobster (Nephrops norvegicus) by functional unit in ICES subarea 7 and also demersal stocks.

### 20.7 Quality of assessment and forecast

Since the benchmark in 2014, UWTV and sampling coverage has been improving in this area. There are now six years of full UWTV survey coverage (2014-2019).

There are several key uncertainties and bias sources in the method used here (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007; WKNEPHBID 2008; SGNEPS 2009; WGNEPS 2014 and ICES, in prep). Ultimately, there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that is more accurate, although no more precise (WKNEPH, 2009). The survey estimates themselves are very precisely estimated (CVs $\sim 5 \%$ ) given the homogeneous distribution of burrow density and the modelling of spatial structuring. The cumulative bias estimates for $\mathrm{FU} 20-21$ are largely based on expert opinion. The precision of these bias corrections cannot yet be characterised, but is likely to be lower than that observed in the survey.

At WGCSE 2018, the group recommended that a review of historical survey data should be undertaken given the large fluctuations observed in the short time-series to date for this survey, that is, to randomly check $20 \%$ of UWTV stations in years 2016 and 2017. This process was conducted in July 2018 during the FU20-21 UWTV survey. The analysis was presented to WGNEPS (ICES, 2018a) and subsequently to the 2019 WGCSE meeting where full details are available in R-markdown (ICES, 2018a; Annex 7). Following this analysis, WGNEPS 2018 recommended that the Manual for the Nephrops Underwater TV Surveys (SISP) (ICES, in prep) to include guidelines on quality control where there are large unexplained fluctuations between abundance estimates from previous years. In that, it is recommended to review $20 \%$ of the survey stations, and when the partial review differs more than $20 \%$ from the survey counts, then a full review of the survey should be considered. This guidelines ensure further quality control of the count data from UWTV surveys.

Sampling of landing and discards for FU20-21 remainslowbut there is a limited number of years for which length-frequency data were available so the three year moving window could only be applied to give two estimates to calculate $\mathrm{F}_{\text {MSY }}$ reference points.

French and Irish trawlers cover different areas and have presented contrasting features over the last decade. The French fleet moved gradually from the "Smalls" Ground (mainly 31E3) to the
"Labadie" (30E2, increase of 28E2 in the early 2010s, although no trend is revealed within FU2021 throughout the overall time-series): in the late 1990s, more than $40 \%$ of French landings were reported from the "Smalls" area whereas by the end of 2000s the contribution of this rectangle became minor (less than 10\%). Irish vessels have increased their production on FU20-21 since the mid-2000s and a gradual expansion towards the southern rectangles is obvious during the recent years (stock annex).

### 20.8 Recommendations for next benchmark

This stock was last benchmarked by WKCELT (ICES, 2014). WGCSE will keep the stock under close review and recommend future benchmark as required.

### 20.9 Management considerations

The indications are the Nephrops in FU20-21 are lightly exploited now relative to the past and recent average landings are broadly sustainable. Overall effort in the fishery has declined to less than $25 \%$ of the peak effort observed in the early 1990s. Harvest rates based on recent landings and UWTV surveys suggest that the HR is low relative to most other Nephrops fisheries.

In recent years, the Irish fishery in the area expanded whereas the French fishery continued to decline. The fishing patterns of the French and Irish fleet are very different with the Irish fleet specialising on Nephrops whereas the French fishery remains more mixed. French Nephrops fisheries in this area are fairly mixed, also catching whiting, cod, megrim, anglerfish and other demersal species (Davie and Lordan, 2011). Nephrops tend to dominate the landings of Irish fisheries in the area but catches are more mixed in the North ( $\sim 50 \%$ Nephrops) and cleaner Nephrops towards the south (~75\% Nephrops) (Gerritsen et al.,2012). The French trawlers showed an overall decline in effort and landings during the last decade, mainly explained by decommissioning schemes associated to constraints linked to fuel prices.

In recent years, several newer vessels specializing in Nephrops fishing have participated in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates.

From 2016, fisheries catching Nephrops in Subarea 7 are covered by the EU landings obligation (EU, 2015). Creel fisheries are exempted from the landings obligation, with a de minimis exemption consisting of a $5 \%$ discard rate by weight for the trawl fishery in 2019 (reduced from $6 \%$ in 2018 and $7 \%$ in both 2016 and 2017). Irish discard survival experiments indicate that the trawl discard survival may be around $64 \%$ (BIM, 2017). As a result, an exemption from the landings obligation based on high survivability has been granted by the European Commission. The average discard rate by weight for FU20-21 over the last three years is $19 \%$. Catch advice and scenarios are provided this year on the assumption that discarding is assumed to continue at the recent average.

UWTV survey coverage has improved. A new survey point available by autumn 2019 provides a more up to date estimate of density and abundance. The use of the most up to date survey information is considered for this stock.

Landings data are adjusted to take into account landings that have been misreported from FU16 since 2011. This adjustment is thought to be reasonably accurate (See Section 19).
ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES division level. Management at the functional unit level could provide controls to ensure effort and catch were in line with resources available.

### 20.10 References

Anon. 2011. Atlas of Demersal Discarding, Scientific Observations and Potential Solutions, Marine Institute, Bord Iascaigh Mhara, September 2011. ISBN 978-1-902895-50-5. 82 pp.

BIM. 2017. Report on Nephrops survivability in the Irish demersal trawl fishery. Bord Iascaigh Mhara, 29 September 2017 .http://www.bim.ie/media/bim/content/publications/fisheries/6882-BIM-nephrops-survival-report-final.pdf.

Charuau A., Morizur Y., Rivoalen J.J. 1982. Survival of discarded Nephrops norvegicus in the Bay of Biscay and in the Celtic Sea. ICES CM 1982/B:13.

EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. Official Journal of the European Union, L 83: 1-17. http://data.europa.eu/eli/reg/2019/472/oj

Davie S., Lordan C. 2011. Definition, dynamics and stability of métiers in the Irish otter trawl fleet. Fish. Res. 111, 145-158. http://dx.doi.org/10.1016/j.fishres.2011.07.005 or http://hdl.handle.net/10793/673.

Gerritsen H.D., Lordan C. 2011. Integrating Vessel Monitoring Systems (VMS) data with daily catch data from logbooks to explore the spatial distribution of catch and effort at high resolution. ICES Journal of Marine Science, 68(1): 245-453.

Gerritsen H.D., Lordan C., Minto C., Kraak S.B.M. 2012. Spatial patterns in the retained catch composition of Irish demersal otter trawlers: High-resolution fisheries data as a management tool. Fisheries Research, Vol. 129-130, pp 127-136. Available from http://www.sciencedirect.com/science/article/pii/S0165783612002032.
ICES. 2009. Report of the Benchmark Workshop on Nephrops assessment (WKNEPH). ICES CM 2009/ACOM:33.

ICES. 2014a. Report of the Benchmark Workshop on Celtic Sea Stocks (WKCELT), 3-7 February 2014, ICES Headquarters, Copenhagen, Denmark. ICES CM $2014 \backslash$ ACOM:42. 194 pp.
ICES. 2014b. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 13-22 May 2014, Copenhagen, Denmark. ICES CM 2014/ACOM:12.

ICES. 2015. Report of the Working Group on Nephrops Surveys (WGNEPS), 10-13 November 2015, Cadiz, Spain. ICES CM 2015/ SSGIEOM:30.52 pp.
ICES. 2015. Report of the International Bottom Trawl Survey Working Group (IBTSWG), 23-27 March 2015, Bergen, Norway. ICES CM 2015/ SSGIEOM:20. 124 pp.

ICES. 2016a. Report of the Working Group on Nephrops Surveys (WGNEPS), 7-8 November 2016. Reykjavík, Iceland. ICES CM 2016/ SSGIEOM:33.67 pp.
ICES. 2016b. Report of the Workshop on NephropsBurrow counting (WKNEPS), 9-11 November 2016. Reykjavík, Iceland. ICES CM 2016/ SSGIEOM:34.65 pp.
ICES. 2016XX. EU request to ICES to provide FMSY ranges for selected stocks in ICES subareas 5 to 10 . In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.2.3.1.

ICES. 2016YY. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.
ICES 2017. Report of the Working Group on NephropsSurveys (WGNEPS). 28 November-1 December 2017, Heraklion, Greece. ICES CM 2017/SSGIEOM: 30. Ref: ACOM/ SCICOM.

ICES. 2018a. Report of the Working Group on Nephrops Surveys (WGNEPS), 6-8 November 2018. Lorient, France. ICES CM 2018/ EOSG:18.227 pp.

ICES. 2018b. Report of the Workshop on Nephrops burrow counting, 2-5 October 2018, Aberdeen, UK. ICES CM 2018/ EOSG:25.47 pp.

ICES. In prep. Manual for the Nephrops Underwater TV Surveys. Series of ICES Survey Protocols (SISP).
Leocádio, A., Weetman, A., and Wieland, K. (Eds). 2018. Using UWTV surveys to assess and advise on Nephrops stocks. ICES Cooperative Research Report No. 340. 49 pp. http://doi.org/10.17895/ices.pub. 4370.

O'Sullivan D., Lordan C., Doyle J., Berry A., Lyons K. 2013. Study of local hydrodynamics and larval dispersal on Nephrops fishing grounds. Irish Fisheries Investigation. No 26: Marine Institute.2014.http://hdl.handle.net/10793/985.

Stokes D., Gerritsen H., O'Hea B., Moore S.J., Dransfeld L. 2014. "Irish Groundfish Survey Cruise Report, 24 September-17 December 2014",FEAS Survey Series; 2014/01.http://hdl.handle.net/10793/1064.

White, J., Aristegui, M., Blaszkowski, M., Fee, D., O’Connor, S., Power, J., Notaro, D., O' Brien and Doyle, J. 2019. The Labadie, Jones and Cockburn Banks Nephrops Grounds (FU20-21) 2019 UWTV Survey Report and catch scenarios for 2020. Marine Institute UWTV Survey report.

Table 23.2.1. Nephrops FU 20-21. Landings in tonnes by country.

| FU 20-21 Landings (t) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | France | Rep. of Ireland | UK | Belgium | Total |
| 1995 | 3419 | 117 | na |  | 3536 |
| 1996 | 2721 | 101 | na |  | 2822 |
| 1997 | 1957 | 81 | na |  | 2038 |
| 1998 | 1583 | 130 | na |  | 1713 |
| 1999 | 1051 | 83 | 18 |  | 1152 |
| 2000 | 1661 | 107 | 10 |  | 1778 |
| 2001 | 1750 | 69 | 14 |  | 1833 |
| 2002 | 2559 | 104 | 11 |  | 2674 |
| 2003 | 2796 | 148 | 9 |  | 2953 |
| 2004 | 2140 | 299 | 4 |  | 2443 |
| 2005 | 2008 | 455 | 6 |  | 2469 |
| 2006 | 2066 | 450 | 7 |  | 2523 |
| 2007 | 1816 | 600 | 3 |  | 2419 |
| 2008 | 2036 | 937 | 7 |  | 2980 |
| 2009 | 1930 | 1202 | 13 |  | 3145 |
| 2010 | 975 | 756 | 62 |  | 1793 |
| 2011 | 566 | 637 | 34 |  | 1237 |
| 2012 | 453 | 708 | 28 |  | 1189 |
| 2013 | 486 | 844 | 57 |  | 1387 |
| 2014 | 465 | 1342 | 29 |  | 1837 |
| 2015 | 355 | 1620 | 141 |  | 2116 |
| 2016 | 477 | 1531 | 445 |  | 2453 |
| 2017 | 341 | 1113 | 395 | 0.2 | 1849 |
| 2018 | 195 | 1197 | 411 | 0.2 | 1803 |

Table 23.2.2. Nephrops FU 20-21. Effort data for the Irish otter trawl Nephrops directed fleet.

| Year | Effort (Kw Days) | Landings (tonnes) |
| :---: | :---: | :---: |
| 1995 | 57 | 104 |
| 1996 | 49 | 74 |
| 1997 | 40 | 59 |
| 1998 | 56 | 102 |
| 1999 | 37 | 48 |
| 2000 | 39 | 62 |
| 2001 | 29 | 45 |
| 2002 | 78 | 165 |
| 2003 | 82 | 86 |
| 2004 | 159 | 164 |
| 2005 | 255 | 360 |
| 2006 | 301 | 348 |
| 2007 | 402 | 512 |
| 2008 | 562 | 920 |
| 2009 | 801 | 1,249 |
| 2010 | 498 | 633 |
| 2011 | 424 | 535 |
| 2012 | 357 | 534 |
| 2013 | 445 | 672 |
| 2014 | 885 | 1,170 |
| 2015 | 1,180 | 1,542 |
| 2016 | 920 | 1,404 |
| 2017 | 704 | 1,004 |
| 2018 | 695 | 1,084 |

Table 23.2.3. Nephrops FU 20-21. Effort data for the French fleet.

| Year | Effort France ('000 hrs) | LPUE France (kg/hr) |
| :---: | :---: | :---: |
| 1983 | 231 | 14 |
| 1984 | 205 | 16 |
| 1985 | 203 | 16 |
| 1986 | 163 | 15 |
| 1987 | 190 | 15 |
| 1988 | 171 | 16 |
| 1989 | 179 | 17 |
| 1990 | 230 | 16 |
| 1991 | 225 | 11 |
| 1992 | 277 | 12 |
| 1993 | 268 | 13 |
| 1994 | 259 | 14 |
| 1995 | 239 | 15 |
| 1996 | 220 | 14 |
| 1997 | 187 | 13 |
| 1998 | 155 | 13 |
| 1999 | 151 | 11 |
| 2000 | 194 | 14 |
| 2001 | 170 | 15 |
| 2002 | 166 | 19 |
| 2003 | 192 | 18 |
| 2004 | 153 | 16 |
| 2005 | 147 | 16 |
| 2006 | 137 | 16 |
| 2007 | 102 | 19 |
| 2008 | 100 | 23 |
| 2009 | 93 | 23 |
| 2010 | 67 | 17 |
| 2011 | 52 | 12 |
| 2012 | 42 | 13 |
| 2013 | 48 | 12 |
| 2014 | 36 | 15 |
| 2015 | 35 | 11 |
| 2016 | 35 | 15 |
| 2017 | 34 | 11 |
| 2018 | 21 | 10 |

Table 23.2.4.a. Nephrops FU 20-21. Sampling levels by Ireland.

| IRELAND |  | Number of Samples |  |  | Numbers Measured |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Quarter | Catch | Discards | Landings | Catch | Discards | Landings |
| 2009 | 2 | 1 |  |  | 489 |  |  |
| 2010 | 2 | 1 |  |  | 461 |  |  |
| 2011 | 2 | 1 |  |  | 270 |  |  |
| 2012 | 1 | 8 | 5 | 1 | 2654 | 2024 | 1747 |
| 2013 | 1 | 1 | 1 |  | 319 | 423 |  |
| 2013 | 2 | 9 | 7 | 1 | 2514 | 2038 | 2187 |
| 2014 | 2 | 2 | 2 |  | 718 | 782 |  |
| 2015 | 1 |  |  | 1 |  |  | 1724 |
| 2015 | 2 | 6 | 6 | 2 | 2714 | 3997 | 3204 |
| 2015 | 3 |  |  | 4 |  |  | 4750 |
| 2015 | 4 | 2 | 2 |  | 650 | 419 |  |
| 2016 | 2 | 8 | 5 | 1 | 2859 | 1485 | 384 |
| 2016 | 4 | 3 | 2 | 4 | 767 | 1678 | 1743 |
| 2017 | 1 | 2 | 1 | 1 | 722 | 297 | 1616 |
| 2017 | 2 | 7 | 4 | 1 | 2813 | 1035 | 365 |
| 2017 | 3 | 3 | 1 |  | 1154 | 296 |  |
| 2017 | 4 | 12 | 7 |  | 3631 | 1983 |  |
| 2018 | 1 | 3 | 3 |  | 987 | 1036 |  |
| 2018 | 2 | 17 | 17 |  | 6691 | 5742 |  |
| 2018 | 3 | 2 |  |  | 389 |  |  |
| 2018 | 4 | 2 | 1 |  | 544 | 369 |  |

Table 23.2.4.b. Nephrops FU 20-21. Sampling levels by France.

| FRANCE <br> Year | Quarter | Number of Samples |  | Numbers Measured |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Landings | Catch | Discards | Landings |
| 2012 | 1 |  | 31 | 9 |  | 391 | 1431 |
| 2012 | 2 |  | 13 | 8 |  | 198 | 1202 |
| 2012 | 3 |  | 47 | 8 |  | 667 | 1155 |
| 2012 | 4 |  | 6 | 6 |  | 16 | 860 |
| 2013 | 1 |  | 0 | 12 |  | 0 | 1362 |
| 2013 | 2 |  | 68 | 72 |  | 1,120 | 3151 |
| 2013 | 3 |  | 16 | 68 |  | 131 | 1917 |
| 2013 | 4 |  | 2 | 14 |  | 12 | 1303 |
| 2014 | 1 |  | 0 | 10 |  | 0 | 1221 |
| 2014 | 2 |  | 40 | 47 |  | 1,127 | 3536 |
| 2014 | 3 |  | 20 | 33 |  | 458 | 1934 |
| 2014 | 4 |  | 0 | 9 |  | 0 | 1360 |
| 2015 | 1 |  | 2 | 14 |  | 60 | 1508 |
| 2015 | 2 |  | 24 | 44 |  | 520 | 3249 |
| 2015 | 3 |  | 1 | 9 |  | 1 | 1366 |
| 2015 | 4 |  | 0 | 9 |  | 0 | 1357 |
| 2016 | 1 |  | 3 | 44 |  | 464 | 3164 |
| 2016 | 2 |  | 4 | 42 |  | 519 | 1263 |
| 2016 | 3 |  | 1 | 25 |  | 217 | 1971 |
| 2016 | 4 |  | 2 | 20 |  | 5 | 1935 |
| 2017 | 1 |  | 3 | 46 |  | 429 | 1659 |
| 2017 | 2 |  | 3 | 80 |  | 852 | 2390 |
| 2017 | 3 |  | 2 | 9 |  | 84 | 344 |
| 2017 | 4 |  | 1 | 23 |  | 307 | 952 |
| 2018 | 1 |  | 8 | 8 |  | 460 | 36 |
| 2018 | 2 |  | 9 | 9 |  | 1190 | 254 |
| 2018 | 3 |  | 30 | 30 |  | 1140 | 105 |
| 2018 | 4 |  | 10 | 10 |  | 149 | 19 |

Table 23.2.5. Nephrops FU 20-21. Sex ratio in the landings by country based on available sampling.

| Ireland |  |  |  |
| :---: | :---: | :---: | :---: |
| Year | Females ('000s) | Males ('000s) | \% Males in Landings |
| 2012 | 1,171 | 25,306 | 96 |
| 2013 | 8,452 | 15,752 | 65 |
| 2014 | 13,630 | 25,467 | 65 |
| 2015 | 8,916 | 39,018 | 81 |
| 2016 | 15,807 | 23,835 | 60 |
| 2017 | 11,836 | 29,183 | 71 |
| 2018 | 15,967 | 28,486 | 64 |
| France |  |  |  |
| 2012 | 1,545 | 9,323 | 86 |
| 2013 | 1,678 | 7,641 | 82 |
| 2014 | 3,292 | 7,316 | 69 |
| 2015 | 1,144 | 6,244 | 85 |
| 2016 | 819 | 8,815 | 91 |
| 2017 | 1,119 | 5,110 | 82 |
| 2018 | 1,863 | 3,605 | 66 |

Table 23.2.6. Nephrops FU 20-21. Landings and discards by number and weight $(t)$, dead discard rate and discard rate by number, discard rate by weight and estimated mean weights (grs) in the landings and discards for France. 25\% discards survival.

| France |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| - ¢ |  |  |  |  |  |  | $\begin{aligned} & \text { a0 } \\ & \stackrel{0}{\overline{0}} \\ & \stackrel{\rightharpoonup}{0} \end{aligned}$ |  |  |  |
|  | millions | millions | millions | \% | \% | \% | tonnes | tonnes | gramme | gramme |
| 2012 | 10.9 | 17.8 | 24.2 | 55.1 | 62.1 | 41.5 | 453 | 322 | 41.7 | 18.1 |
| 2013 | 9.3 | 10.0 | 16.9 | 44.7 | 51.9 | 26.6 | 486 | 176 | 52.2 | 17.6 |
| 2014 | 10.6 | 37.0 | 38.4 | 72.4 | 77.7 | 55.8 | 465 | 588 | 43.8 | 15.9 |
| 2015 | 7.4 | 7.7 | 13.2 | 43.9 | 51.1 | 31.7 | 355 | 165 | 48.1 | 21.4 |
| 2016 | 9.6 | 3.2 | 12.0 | 19.7 | 24.7 | 16.2 | 477 | 92 | 49.5 | 29.1 |
| 2017 | 6.2 | 5.9 | 10.7 | 41.6 | 48.7 | 26.2 | 341 | 121 | 54.8 | 20.5 |
| 2018 | 5.5 | 4.7 | 9.0 | 39.0 | 46.1 | 32.3 | 195 | 93 | 35.6 | 19.9 |

Table 23.2.7. Nephrops FU 20-21. Landings and discards by number and weight ( $\mathbf{t}$ ), dead discard rate and discard rate by number, discard rate by weight and estimated mean weights (grs) in the landings and discards for Ireland. 25\% discards survival.

| Ireland |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\stackrel{\text { ¢ }}{\text { ¢ }}$ |  |  |  |  |  |  | $\begin{aligned} & \text { M } \\ & \stackrel{0}{0} \\ & \stackrel{0}{0} \end{aligned}$ |  |  |  |
|  | millions | millions | millions | \% | \% | \% | tonnes | tonnes | gramme | gramme |
| 2012 | 26.5 | 17.5 | 39.6 | 33.1 | 39.7 | 22.6 | 708 | 207 | 26.7 | 11.9 |
| 2013 | 24.2 | 8.3 | 30.5 | 20.5 | 25.6 | 14.0 | 844 | 137 | 34.9 | 16.4 |
| 2014 | 39.1 | 17.6 | 52.3 | 25.3 | 31.1 | 14.8 | 1342 | 233 | 34.3 | 13.3 |
| 2015 | 47.9 | 18.6 | 61.9 | 22.5 | 27.9 | 13.3 | 1620 | 248 | 33.8 | 13.4 |
| 2016 | 39.6 | 27.5 | 60.3 | 34.2 | 41.0 | 26.9 | 1531 | 564 | 38.6 | 20.5 |
| 2017 | 41.0 | 9.2 | 47.9 | 14.4 | 18.4 | 9.7 | 1113 | 120 | 27.1 | 13.0 |
| 2018 | 44.5 | 11.9 | 53.4 | 16.8 | 21.2 | 14.4 | 1197 | 201 | 26.9 | 2018 |

Table 23.2.8. Nephrops FU 20-21. Landings and discards by number and weight $(t)$, dead discard rate and discard rate by number, discard rate by weight and estimated mean weights (grs) in the landings combined by both countries based on available sampling and scaled to international landings. $\mathbf{2 5 \%}$ discards survival.

Combined and scaled to the international landings

| $\begin{aligned} & \bar{\varpi} \\ & \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions | millions | millions | \% | \% | \% | tonnes | tonnes | gramme | gramme |
| 2012 | 38.2 | 36.1 | 65.3 | 41.4 | 48.5 | 31.3 | 1,189 | 542 | 31.1 | 15.0 |
| 2013 | 34.8 | 19.2 | 49.2 | 29.3 | 35.6 | 19.1 | 1,387 | 327 | 39.9 | 17.0 |
| 2014 | 50.6 | 55.5 | 92.2 | 45.2 | 52.3 | 31.2 | 1,836 | 834 | 36.3 | 15.0 |
| 2015 | 59.4 | 28.1 | 80.5 | 26.2 | 32.2 | 17.3 | 2,116 | 442 | 35.7 | 15.7 |
| 2016 | 60.2 | 37.5 | 88.3 | 31.8 | 38.4 | 24.6 | 2,453 | 801 | 40.7 | 21.4 |
| 2017 | 60.1 | 19.2 | 74.5 | 19.4 | 24.3 | 14.2 | 1,849 | 306 | 30.8 | 15.9 |
| 2018 | 64.7 | 21.5 | 80.8 | 20.0 | 25.0 | 17.5 | 1803 | 381 | 27.9 | 17.7 |

Table 23.2.9. Nephrops FU 20-21. Results summary table for geo-statistical analysis of UWTV survey.

| Ground | Year | Number of stations | Mean Density adjusted (burrows/m²) | Domain <br> Area (km²) | Geostatistical Abundance Estimate adjusted (millions burrows) | CV on Burrow estimate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FU20-21 | 2006 | 9 | 0.44 |  | nr | nr |
|  | 2012 | 54 | 0.57 |  | nr | nr |
|  | 2013 | 55 | 0.16 | 5,701 | 942 | 3\% |
|  | 2013* | 55 |  | 10,014 | 1624 |  |
|  | 2014 | 98 | 0.19 | 10,014 | 2051 | 3\% |
|  | 2015 | 96 | 0.2 | 10,014 | 2003 | 3\% |
|  | 2016 | 93 | 0.18 | 10,014 | 1879 | 5\% |
|  | 2017 | 86 | 0.44 | 10,014 | 4428 | 4\% |
|  | 2018 | 96 | 0.27 | 10,014 | 2721 | 4\% |
|  | 2019 | 95 | 0.06 | 10,014 | 617 | 5\% |

[^12]Table 23.3.1. Nephrops FU 20-21. Short-term catch options prediction inputs and recent estimates of mean weight in landings and harvest rates. Cells in bold indicates inputs to catch option calculations.




Figure 23.1.1. Nephrops FU 20-21. Number of Irish vessels reporting landings $>10 \mathrm{t}$ by year.


Figure 23.1.2. Nephrops FU 20-21. Combined box and kite plot of vessel power on the FU20-21 grounds by year. The blue line indicates the mean.


Figure 23.2.1. Nephrops FU 20-21. Landings in tonnes by country.


Figure 23.2.2. Nephrops FU 20-21. Effort data (Kw days) for the Irish otter trawl Nephrops directed fleet.


Figure 23.2.3. Nephrops FU 20-21. Commercial length-frequency distribution by country. Minimum conservation reference size of 25 CL mm (European MCR) and 35 CL mm (French MLS) display ed.


Figure 23.2.4. Nephrops FU 20-21. Annual mean weights (gr) in the landings (blue line) and discards (red line) by country and combined scaled to international landings.


Figure 23.2.5. Nephrops FU 20-21. Annual discard ogive derived from Irish sampling. Minimum landing size of 25 CL mm (European MCR) as black line.


Figure 23.2.6. Nephrops FU 20-21. Contour plots of krigged density estimates for the UWTV surveys from 2013 to 2019.


Figure 23.2.7. Nephrops FU 20-21. Time-series of abundance estimates for FU20-21 (error bars indicate 95\% confidence intervals).

## Length frequencies for IGFS Survey Catches: <br> Nephrops in FU2021



Figure 23.2.8. Nephrops FU 20-21. Mean size trends for catches by sex from the IBTS-IGFS Irish survey in the Celtic Sea.

Length frequencies for catch (dotted) and landed(solid):
Nephrops in FU2021


Figure 23.2.9. Nephrops FU 20-21. Mean size trends for catches by sex from the IBTS-EVHOE French survey in the Celtic Sea. No survey data available for 2017.


Figure 23.3.1. Nephrops FU 20-21. Harvest ratio (\% dead removed / UWTV abundance). The dashed and solid lines are the MSY proxy and the harvest rate respectively.

# 21 Norway lobster (Nephrops norvegicus) in divisions 7.g and 7.f, Functional Unit 22 (Celtic Sea, Bristol Channel) 

## Type of assessment in 2019

UWTV based assessment using WKNEPH 2009 protocol as described in the stock annex. The TV survey is due to be repeated in the summer 2019, and the new survey will form the basis of advice for this stock in the autumn.

## ICES advice applicable to 2018

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2014-2016, catches in 2018 should be no more than 4322 tonnes.

To ensure that the stock in functional unit (FU) 22 is exploited sustainably, management should be implemented at the functional unit level."

## ICES advice applicable to 2019

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of 2015-2017, catches in 2019 should be no more than 2084 tonnes.

To ensure that the stock in Functional Unit 22 is exploited sustainably, management should be implemented at the functional unit level."

### 24.1 General

## Stock description and management units

The Smalls Nephrops stock (FU22) covers ICES rectangles 31-32E3,31-32E4 within 7.f.g. It is included in the whole ICES Area 7 together with Irish Sea East and West [FU14, FU15], Porcupine Bank [FU16], Aran Grounds [FU17], northwest Irish Coast [FU18], southeast and southwest Irish Coast [FU19], NW Labadie, Baltimore and Galley [FU20-21], Jones and Cockburn [FU21].

Historically FU20-22 has covered an amalgamation of several spatially distinct mud patches;FU 20 NW Labadie, Baltimore and Galley, FU 21 Jones and Cockburn and FU22 the Smalls. There is no evidence that the whole exploited area belongs to the same stock, or that there are several patches linked in meta-population sense. WGCSE 2013 recommended that FU20-22 should be split into FU20-21 and FU22 for the purposes of assessment and advice provision. The map below shows FU22 assessment area (blue) and TAC area (red). There is no evidence that the individual functional units belong to the same stock. See Section 18 for details on Nephrops in Subarea 7 general section.


## Ecosystem aspects

This section is detailed in stock annex.

## Fishery description

Ireland, France and the UK are the main countries involved in the FU22 Nephrops fishery. In the early 2000s, the Republic of Ireland fleet had on average over $70 \%$ of the landings and this has increased to over $90 \%$ from this FU in recent times. A description of this fleet is given in the stock annex. The time-series of numbers of vessels is updated in Figure 24.1.1. The numbers of vessels has been decreasing in recent years where the highest number was recorded in 2016. The time-series of vessel power is shown as a box and kite plot in Figure 24.1.2.
Irish landings from this FU come mainly from ICES statistical rectangle 31E3. The fishery on the Smalls grounds operates throughout the year, weather permitting with a seasonal trend.
French trawlers targeting Nephrops in the Celtic Sea operate mainly in FU20-21. In the early 2000 s, French fleet had on average $30 \%$ of the landings from FU22 where this has decreased to $\sim 2 \%$ in recent times. $80-90 \%$ of the FU22 French landings come from ICES statistical rectangle 31 E 3.

UK fleet had on average $\sim 10 \%$ of the landings in recent year and is mainly UK-Northern Irish vessels in this fishery.

## Fishery in 2018

In 2018, 74 Irish vessels reported landings from FU22. Of these, 57 vessels reported landings in excess of 10 t . Vessels $>18 \mathrm{~m}$ account for $90 \%$ of the landings in 2018. In recent years, several newer vessels specializing in Nephrops fishing have participated periodically in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates. In 2018, 15 French trawlers reported landings for FU22. French vessels switch between FU20-21 and FU22. In 2018, 34 UK vessels reported landings for this FU.
The French minimum mesh size of codend was set at 100 mm since January 2000 , the majority of Irish landings are from vessels with $80-99 \mathrm{~mm}$ codend mesh.

## Information from stakeholders

None presented.

### 24.2 Data

## InterCatch

Data were available in InterCatch and used for catch data only. French catch data provided directly by the national expert.

## Landings

The reported landings time-series by country is shown in Figure 24.2.1 and Table 24.2.1. The reported Irish landings from FU22 have increased since 2000. In 2018, the landings decreased from 2017 by $42 \%$ to approximately 1640 t . French landings have gradually decreased since the early 2000s to the present to the lowest level (3t). Reported landings from the UK have fluctuated with a recent (three year) increasing trend coming mainly from the Northern Irish fleet. In 2018, Northern Ireland had thehighest landings at 258 t followed by England and Wales reporting 48 t and 25 t from Scotland. Belgium reported minimal landings in 2018.

## Effort

In line with WGCSE 2015 recommendation effort is reported in Kwdays and LPUE reported in $\mathrm{t} / \mathrm{Kwdays}$ in the knowledge that the trend is likely to be a biased underestimate because it is not adjusted for efficiency or behavioural changes. The effort series is based on the same criteria for FU15, 16, 17, 22 and 20-21 ( $30 \%$ landings threshold) and will be contingent on the accuracy of landings data reported in logbooks. Effort data are available for the Irish Nephrops directed fleet in FU22 from 1995-2018. The time-series of effort and LPUE is updated in Figure 24.2.2 and Table 24.2.2.

Effort shows an increasing trend since the early 2000s (Table 24.2.2. and Figure 24.2.2) with a decrease in 2018.

## Sampling levels

A dedicated sampling of landings and discards began in 2003 by Ireland. Sampling levels in 2018 were good and comparable to levels in 2017 (Figure 24.2.3).

## Sampling and Raising Procedure Review

The national sample raising procedures for FU22 were reviewed and fully documented through an $r$ markdown document (Annex 3, ICES, 2018 and stock annex). Annual discard ogives are calculated and are applied to quarterly length distributions and then raised to total quarterly landings before aggregation. A further raising procedure is applied to raise the annual sampled Irish data where these address quarters with missing length samples. Next, the international raising factor is applied. This raising procedure is used to assess this stock and to calculate mean weights, sex ratio and discard rates as inputs for catch scenarios and advice.

## Commercial length-frequency distributions

The Irish sampling programme started in 2003 and since then coverage and intensity have been very good covering the seasonal trend of the fishery. The mean size of Nephrops in Irish landings has remained stable for both sexes. The mean size of Nephrops in the catch has remained relatively stable since 2005 (Figure 24.2.5). There is decrease in mean size in the catches in 2007 to 2009 for both sexes, which is linked to the recruitment signal picked up by both the 2006 UWTV and Irish groundfish survey.

## Sex ratio

The sex ratio by year is shown in Figure 24.2.6. This shows some fluctuations over time. The sex ratio has a distinct seasonal pattern (Figure 24.2.7) with lowest males proportions in the samples in May and June. Males dominate the catches in the autumn and winter.

## Mean weight explorations

Explorations of the mean weight in the catch samples by sex shows a strong cyclical pattern in the females (Figure 24.2.7). This corresponds with the emergence of mature females from the burrows to mate in summer. There is a decrease in mean weight in 2007 to 2009 for both sexes which is linked to the recruitment signal picked up by both the beam trawl during the UWTV in 2006 and 2007 Doyle et al, and Irish groundfish survey (Figure 24.2.11). The annual mean weight estimate for landings and discards is shown in Figure 24.2.8. The mean weight estimates in the landings show a slight decrease but are stable in recent four years.

## Discarding

Since 2003 discard rates have been estimated using unsorted catch and discards sampling. This involves unsorted catch and discard samples being provided by vessels or collected by observers at sea on discard trips. The catch sample is partitioned into landings and discards using an onboard discard selection ogive derived for the discard samples. Sampling effort is stratified monthly, but annual aggregations are used to derive length distributions and selection ogives. The length-weight regression parameters given in the stock annex are used to calculate sampled weights and appropriate quarterly raising factors. The sampling intensity and coverage has varied over the time-series, but overall has been good.

Discard rates range between $9-39 \%$ of total catch by weight and $15-52 \%$ of total catch by number (Table 24.2.4). Discard rate of females tends to be higher due to the smaller average size and market reasons. There is no information on discard survival rate in this fishery. $25 \%$ is assumed in line with other Nephrops stocks in the Celtic Sea (Charuau et al., 1982). Highest discard rates were observed in 2007 as a result of the recruitment into the fishery in 2006.

Gear selectivity trials by Bord Iascaigh Mhara (BIM, 2017) reported a $64 \%$ survivor rate for Nephrops caught in a trawl with a SELTRA selectivity device in the outer Galway Bay area.

Table 24.3.1 gives weights, numbers and mean weights of the landings and discard raised internationally according to the stock annex.

## Surveys

## Abundance indices from UWTV surveys

The methods used during the survey were similar to those employed for UWTV surveys of Nephrops stocks around Ireland and elsewhere and are documented by WKNEPHTV (ICES, 2007), SGNEPS (ICES, 2009; 2010; 2012), WGNEPS (ICES, 2013; 2014; 2015; 2016a, 2017a, 2018a), WKNEPS (2016b, 2018b) and Leocádio et al. (2018). SGNEPS 2012 (ICES, 2012) recommended that a CV (or relative standard error) of $<20 \%$ is an acceptable precision level for UWTV survey estimates of abundance. This allowed sampling intensity to be reduced from around 90 stations in the past to around 41 on the Smalls grounds in 2019, which allowed survey coverage of other FUs. A randomised isometric grid design was employed with UWTV stations at 4.5 nmi intervals, whereas previously a 3.0 nmi square grid was used. Operational details of the 2019 UWTV survey are available (Doyle et al., 2019).
Seven stations in FU22 were not surveyed successfully in 2015 due to very poor visibility conditions encountered as a result of strong tides. WKCELT 2014 concluded that WGCSE or WGNEPS should make recommendations on the most appropriate fill in procedure to be adopted in cases when stations could not be surveyed. WGCSE 2015 agreed the following procedure for this case: Two buffer zones of 1 nmi and 2 nmi distance were generated around the missing stations. The counts and mean of historic density estimates within the 1 and 2 nmi buffers were calculated. The standard kriging procedure was carried out and summary results were computed for the 1 and 2 nmi "fill-ins". Finally the mean of historic densities within 2 nmi buffer of the planned stations were used in the calculation of the 2015 abundance.

The blanked krigged contour plot and posted point density data are shown in Figure 24.2.9. The krigged contours correspond very well to the observed data. In general, the densities are higher in the central area of the ground with a localised hotspot centrally and also in the southwestern leg. Densities and abundance have remained stable in the time-series with the exception of the first year and 2017, which were the highest in the series. The 2019 mean density 0.40 burrows $/ \mathrm{m}^{2}$ is approximately $30 \%$ increase compared with 0.31 burrows $/ \mathrm{m}^{2}$ in 2018. The summary statistics from this geostatistical analysis are given in Table 24.2.5 and plotted in Figure 24.2.10.

The 2019 estimate of 1121 million burrows is above the MSY Btrigger ( 990 million). The estimation variance of the survey as calculated by EVA is very low (CVs in the order $<9 \%$ ).

## Groundfish survey data

The Irish groundfish survey (IGFS-WIBTS-Q4) has been carried out since 2003 (Stokes etal., 2014; ICES, 2017b). This provides information on length-frequency compositions, mean size in the catches, CPUE of Nephrops in FU22. The mean size of the catches is stable over the time-series except in 2006 and 2008, which signals recruitment into the fishery in 2006 and 2007 (Figure 24.2.11). This signal of recruitment was also picked up from the length frequency distributions from beam trawls during the 2006 UWTV survey (Doyle et al., 2019). The groundfish survey provides a useful indicator of recruitment in this FU.

### 24.3 Assessment

## Comparison with previous assessments

The WGCSE2019 carried out an UWTV based assessment for this stock. The methods used were in line with WKNEPH (ICES, 2009) and the approach taken for other Nephrops stocks in areas 6 and 7 by WGCSE.

## State of the stock

UWTV abundance estimates suggest that the stock size shows a recent declining trend with an increase in 2019. The 2019 estimate is above the MSY $B_{\text {trigger }}(990$ million). The 2019 estimate ( 1121 million) is below the average of the series (geomean [2006-2019]: 1223 million).

Harvest rate is calculated as (landings + dead discards)/(abundance estimate). Table 24.3.1 and Figure 24.3 .1 summarize recent harvest rates. The recent fluctuations in harvest rates are a combination of large changes in recent UWTV abundance and variable landings. The current harvest rate is $13.8 \%$, which is just above $\mathrm{F}_{\text {msy }}$.

### 24.4 Catch scenarios table

Catch scenario table inputs and historical estimates of mean weight in landings and harvest ratios are presented in Table 24.3.1 and summarised below.

Since 2003, mean weight in the landings has varied between 18-27 gr (Figure 24.2.8). In previous assessments the long-term average (rather than a three-year average) was used as input for the mean weight in landings and discards in the calculation of catch scenarios, to account for interannual variation. In 2019, the last three year average was considered to be more appropriate given that average weights have been stable at a lower level since 2015.
Three year average (2016-2018) of proportion of removals retained was used as is standard for other Nephrops stocks. The estimate harvest rate has also varied a lot, $6-27 \%$ with 2007 being the highest observed (Figure 24.3.1). This is a result of recruitment into the fishery in 2006 and 2007.

The basis for the catch scenarios.

| Variable | Value | Notes |
| :--- | :---: | :--- |
| Stock abundance | 1121 million <br> individuals | UWTV survey 2019 |
| Mean weight in <br> wanted catch | 20.9 g | Average 2016-2018 |
| Mean weight in <br> unwanted catch | 10.6 g | Average 2016-2018 |
| Unwanted catch <br> rate | $22.8 \%$ | Average 2016-2018 (by number). Calculated as discards divided by landings + dis- <br> cards. |
| Discard survival <br> rate | $25.0 \%$ | Only applies in scenarios where discarding is allowed. <br> Dead Unwanted <br> rate <br> $18.2 \%$ |

A prediction of landings for the FU22 using the approach agreed procedure proposed at WKNEPH 2009 and outlined in the stock annex will be made on the basis of the 2019 UWTV survey. This will be presented in October 2020 for the provision of advice.

### 24.5 Reference points

New reference points were derived by WKMSYRef4 (ICES, 2016XX, 2016YY) for FU22. These were updated on the basis of an average of estimated FMSY proxy harvest rates over a period of years, this corresponds more closely to the methodology for finfish. In cases where there is a clear trend in the values, a five-year average was chosen. Similarly, the five-year average of the F at $95 \%$ of the YPR obtained at the FMSY proxy reference point was proposed as the Fmsy lower bound and the five-year average of the $F$ above $F_{\text {max }}$ that leads to YPR of $95 \%$ of the maximum as the upper bound. Using an average value also has the advantage of reducing the effect of any unusually high or low estimates of the FMSY proxy, which occasionally appear.

This stock previously did not have MSY Btrigger specified; the time-series and range of indicator biomass is alsolimited such that direct use of Bloss is considered too close to equilibrium biomass. The workshop proposed to use the $5 \%$ interval on the probability distribution of indicator biomass assuming a normal distribution, which is analogous to the $5 \%$ on BMSY proposed for finfish stocks assuming these Nephrops FU have been exploited at a rate close to near HRnsy. The MSY $B_{\text {trigger }}$ for FU22 is 987 million individuals rounded to 990 million.

| Stock code | MSY Flower* | F $_{\text {MSY }}{ }^{*}$ | MSY Fupper* with AR | MSY $B_{\text {triger }}$ | MSY Fupper* with no AR |
| :--- | :---: | :---: | :---: | :---: | :---: |
| nep-22 | $10.2 \%$ | $12.8 \%$ | $12.8 \%$ | $990^{* * *}$ | $12.8 \%$ |

* Harvest rate (HR).
*** Abundance in millions.


### 24.6 Management strategies

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to Norway lobster (Nephrops norvegicus) by functional unit in ICES Subarea 7, and also to demersal stocks.

### 24.7 Quality of assessment and forecast

Since 2006, a dedicated annual UWTV survey has provided abundance estimates for FU22 with high precision. There are several key uncertainties and bias sources in the method used here (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007; WKNEPHBID 2008; SGNEPS 2009; WGNEPS 2016, WGNEPS 2018b). Ultimately, there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that is more accurate, although no more precise (WKNEPH 2009). The survey estimates themselves are very precisely estimated (CVs 2-9\%) given the homogeneous distribution of burrow density and the modelling of spatial structuring. The cumulative bias estimates for FU22 are largely based on expert opinion. The precision of these bias corrections cannot yet be characterised, but is likely to be lower than that observed in the survey.

In 2015, there is added uncertainty, not accounted for in the model or CV estimate, because $17 \%$ of the planned TV stations could not be successfully surveyed due to poor visibility on the seabed. However, the spatial distributions of densities have been fairly consistent over time and the overall density has also been relatively stable. The fill in procedure used to generate density estimates for the seven missing stations should be a good approximation.

A review of sampling and raising procedures was presented to WGCSE 2018, and is accepted as the current method to calculate the fishery dependant inputs FU22 (Annex 3, ICES, 2018 and stock annex).

In the provision of catch scenarios based on the absolute survey estimates, additional uncertainties related to mean weight in the landings and the discard rates also arise. For FU22 deterministic estimates of the mean weight in the landings and discard rates for 2003-2018have been used previously by the WG to account for the variability in these over time. This variability has occurred when large recruitments are observed in the stock as was the case in 2006 and 2007. The mean weights have been stable since 2015, and hence the recent 3 year average is used in the catch scenarios for 2020.

From 2016, fisheries catching Nephrops in Subarea 7 are covered by the EU landings obligation with several exemptions (EU, 2015). The average discard rate by weight for FU22 over the last three years is $13 \%$. Irish discard survival experiments indicate that the trawl discard survival may be around $64 \%$ (BIM, 2017). As a result, an exemption from the landings obligation based on high survivability has been granted by the European Commission. Catch advice and scenarios are provided this year on the assumption that discarding is assumed to continue at the recent average.

Landings data are adjusted to take into account landings that have been misreported from FU16 since 2011. This adjustment is thought to be reasonably accurate (See Section 20).

Sampling and discard estimates have improved over the time-series.

### 24.8 Recommendation for next benchmark

This stock has not been formally benchmarked by ICES although the approach used has.WGCSE recommends that the issue list below can be addressed through an inter-bench process:

- The biological parameters used as inputs to the SCA should be reconsidered; growth parameters, length-at-maturity and natural mortality.
- The historical time-series of landings and effort by rectangle should be disaggregated and options for standardisation of LPUE investigated.
- Historical sampling and groundfish survey data in this FU should also be disaggregated as far as possible back in time and investigated for useful trends and signals.


### 24.9 Management considerations

The trends from the fishery (landings, effort, mean size, etc.) appear to be relatively stable. The UWTV abundance and mean density estimates show some fluctuations in burrow abundance in the recent five years although it is stableover the time-series. There are fluctuations in the harvest rates, which are related to the recent large changes in stock abundance. Recent harvest rates for the FU22 Smalls fluctuate and suggest the stock is exploited above Fms.

This up to date survey information will be used to generate catch options and the provision of advice in October 2020.

In recent years, several newer vessels specializing in Nephrops fishing have participated in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates. There have been concerns that effort could be displaced towards the Smalls and other Nephrops grounds due to effort controls in 7.a and 6.a. This has not happened to date and the 2014 effort was just below the recent average in the time-series.

There has been a trend for Irish vessels ( $>18 \mathrm{~m}$ ) to switch to multi (quad) rig trawls. Provisional data suggest a $\sim 30 \%$ increase in Nephrops catch rates and a reduction in fish bycatch of $\sim 30 \%$ due to the lower headline height.

Nephrops fisheries in the Smalls have non-Nephrops bycatch composition. Cod, whiting and to a lesser extent haddock are the main bycatch species (Davie and Lordan, 2011). A target whiting fishery also overlaps with the Nephrops fishery in this area but this has negligible bycatch of Nephrops.

### 24.10 References

Anon. 2011. Atlas of Demersal Discarding, Scientific Observations and Potential Solutions, Marine Institute, Bord Iascaigh Mhara, September 2011. ISBN 978-1-902895-50-5. 82 pp.
BIM. 2017. Report on Nephrops survivability in the Irish demersal trawl fishery. Bord Iascaigh Mhara, 29 September 2017 .http://www.bim.ie/media/bim/content/publications/fisheries/6882-BIM-nephrops-survival-report-final.pdf.
Charuau A., Morizur Y., Rivoalen J.J. 1982. Survival of discarded Nephrops norvegicus in the Bay of Biscay and in the Celtic Sea. ICES CM 1982/B:13.

Davie S., Lordan C. 2011. Definition, dynamics and stability of métiers in the Irish otter trawl fleet. Fish. Res. 111, 145-158.http://dx.doi.org/10.1016/j.fishres.2011.07.005 or http://hdl.handle.net/10793/673.
Doyle, J., O'Brien, S., Fitzgerald, R., Vacherot, JP., Sugrue, S., and Quinn M. 2019. The "Smalls" Nephrops Grounds (FU22) 2019 UWTV Survey Report and catch scenarios for 2020. Marine Institute UWTV Survey report.

EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. Official Journal of the European Union, L 83: 1-17. http://data.europa.eu/eli/reg/2019/472/oj
ICES. 2007. Report of the Workshop on the use of UWTV surveys for determining abundance in Nephrops stocks throughout European waters (WKNEPHTV). ICES CM: 2007/ACFM:14.
ICES. 2008. Report of the Workshop and training course on Nephrops burrow identification (WKNEPHBID). ICES CM 2008/LRC:03.

ICES. 2009. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 2009/LRC: 15, pp 52.
ICES. 2009. Report of the Benchmark Workshop on Nephrops assessment (WKNEPH). ICES CM 2009/ACOM:33.
ICES. 2010. Report of the Study Group on Nephrops Surveys (SGNEPS), 9-11 November 2010, Lisbon, Portugal. ICES CM 2010/SSGESST:22.

ICES. 2012. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 012/SSGESST:19.
ICES. 2014. Report of the Working Group on Nephrops Surveys (WGNEPS), 4-6 November 2014, Lisbon, Portugal. ICES CM 2014/SSGESST:20.57 pp.
ICES. 2015. Report of the Working Group on Nephrops Surveys (WGNEPS), 10-13 November 2015, Cadiz, Spain. ICES CM 2015/ SSGIEOM:30.52 pp.
ICES. 2016b. Report of the Workshop on NephropsBurrow counting (WKNEPS), 9-11 November 2016. Reykjavík, Iceland. ICES CM 2016/ SSGIEOM:34.65 pp.
ICES. 2016XX. EU request to ICES to provide Fmsy ranges for selected stocks in ICES subareas 5 to 10 . In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.2.3.1.

ICES. 2016YY. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.
ICES. 2017a. Interim Report of the Working Group on Nephrops Surveys (WGNEPS), 28 November-1 December 2017, Heraklion, Greece. ICES CM 2017/ SSGIEOM:19 pp.
ICES. 2017b. Interim Report of the International Bottom Trawl Survey Working Group (IBTSWG), 27-31 March 2017, Copenhagen, Denmark. ICES CM 2017/ SSGIEOM:01. 343 pp.
ICES. 2018a. Report of the Working Group on Nephrops Surveys (WGNEPS), 6-8 November 2018. Lorient, France. ICES CM 2018/ EOSG:18. 227 pp.
ICES. 2018b. Report of the Workshop on Nephrops burrow counting (WKNEPS), 2-5 October 2018, Aberdeen, UK. ICES CM 2018/ EOSG:25.47 pp.
ICES .In prep. Manual for the Nephrops Underwater TV Surveys. Series of ICES Survey Protocols (SISP).
Leocádio, A., Weetman, A., and Wieland, K. (Eds). 2018. Using UWTV surveys to assess and advise on Nephrops stocks. ICES Cooperative Research Report No. 340. 49 pp. http://doi.org/10.17895/ices.pub. 4370.
O’Brien, S., Blaszkowski, M., Butler, R., Fee, D., Hernon, P., Santana, C., Lordan, C. and Doyle, J. 2017. The "Smalls" Nephrops Grounds (FU22) 2017 UWTV Survey Report and catch options for 2018. Marine Institute UWTV Survey report.

Marrs S.J., Atkinson R.J.A., Smith C.J., Hills J.M. 1996. Calibration of the towed underwater TV technique for use in stock assessment of Nephrops norvegicus. Reference no. 94/069 (Study Project in support of the Common Fisheries Policy XIV/1810/C1/94, call for proposals 94/C 144/04).

Stokes, D., Gerritsen. H., O'Hea, B., Moore, S.J. and Dransfeld, L. 2014. "Irish Groundfish Survey Cruise Report, 24 September-17 December 2014", FEAS Survey Series; 2014/01. http://hdl.handle.net/10793/1064.

Table 24.2.1. Nephrops in FU22 (Smalls Grounds). Landings in tonnes by country.

| FU 22 Landings (t) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | France | Rep. of Ireland | UK | Belgium | Total |
| 1999 | 1034 | 741 | 0 |  | 1775 |
| 2000 | 1192 | 1687 | 11 |  | 2890 |
| 2001 | 882 | 2054 | 2 |  | 2938 |
| 2002 | 598 | 1392 | 3 |  | 1993 |
| 2003 | 799 | 1257 | 10 |  | 2065 |
| 2004 | 454 | 1349 | 26 |  | 1828 |
| 2005 | 478 | 1987 | 68 |  | 2533 |
| 2006 | 293 | 1442 | 19 | 7 | 1761 |
| 2007 | 216 | 2716 | 13 | 5 | 2950 |
| 2008 | 301 | 2539 | 241 | 9 | 3090 |
| 2009 | 258 | 1609 | 306 | 12 | 2185 |
| 2010 | 129 | 2219 | 351 | 15 | 2714 |
| 2011 | 64 | 1521 | 44 | 7 | 1636 |
| 2012 | 65 | 2506 | 41 | 6 | 2618 |
| 2013 | 83 | 2054 | 107 | 12 | 2257 |
| 2014 | 29 | 2428 | 61 | 8 | 2526 |
| 2015 | 9 | 2215 | 121 | 5 | 2350 |
| 2016 | 5 | 2967 | 354 | 3 | 3329 |
| 2017 | 7 | 2815 | 737 | 1 | 3560 |
| 2018 | 3 | 1639 | 331 | 1 | 1974 |

Table 24.2.2. Nephrops in FU22 (Smalls Grounds). Effort data for the Irish otter trawl Nephrops directed fleet.

| Year | Effort ( '000s Kw Days) | Landings (tonnes) | LPUE (t/KwDays) |
| :---: | :---: | :---: | :---: |
| 1995 | 552 | 1226 | 2.2 |
| 1996 | 412 | 1010 | 2.5 |
| 1997 | 474 | 1096 | 2.3 |
| 1998 | 524 | 1353 | 2.6 |
| 1999 | 292 | 620 | 2.1 |
| 2000 | 586 | 1335 | 2.3 |
| 2001 | 789 | 1964 | 2.5 |
| 2002 | 615 | 1298 | 2.1 |
| 2003 | 639 | 1000 | 1.6 |
| 2004 | 620 | 981 | 1.6 |
| 2005 | 986 | 1882 | 1.9 |
| 2006 | 855 | 1374 | 1.6 |
| 2007 | 1131 | 2677 | 2.4 |
| 2008 | 1047 | 2501 | 2.4 |
| 2009 | 702 | 1605 | 2.3 |
| 2010 | 962 | 2198 | 2.3 |
| 2011 | 724 | 1497 | 2.1 |
| 2012 | 970 | 2260 | 2.3 |
| 2013 | 902 | 1849 | 2.0 |
| 2014 | 915 | 2182 | 2.4 |
| 2015 | 971 | 2076 | 2.1 |
| 2016 | 1270 | 2761 | 2.2 |
| 2017 | 1229 | 2712 | 2.2 |
| 2018 | 748 | 1509 | 2.0 |

Table 24.2.4. Nephrops in FU22 (Smalls Grounds). Landings and discards weight and numbers by year.

| Year | Landings (t) | Discards (t) | Landings ('000s numbers) | Discards ('000s numbers) | \% Discard by weight | \% Discard by number |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 1257 | 438 | 57.9 | 41.1 | 25.8 | 41.5 |
| 2004 | 1349 | 149 | 52.1 | 9.7 | 9.9 | 15.6 |
| 2005 | 1987 | 1292 | 93.6 | 100.9 | 39.4 | 51.9 |
| 2006 | 1442 | 372 | 82.0 | 37.0 | 20.5 | 31.1 |
| 2007 | 2716 | 1755 | 152.1 | 166.5 | 39.3 | 52.3 |
| 2008 | 2539 | 237 | 118.0 | 21.4 | 8.5 | 15.3 |
| 2009 | 1609 | 274 | 67.7 | 24.3 | 14.5 | 26.4 |
| 2010 | 2219 | 520 | 99.6 | 36.4 | 19.0 | 26.8 |
| 2011 | 1521 | 183 | 55.7 | 12.2 | 10.7 | 18.0 |
| 2012 | 2506 | 332 | 115.2 | 30.0 | 11.7 | 20.7 |
| 2013 | 2054 | 452 | 85.1 | 36.5 | 18.1 | 30.0 |
| 2014 | 2428 | 442 | 96.3 | 32.1 | 15.4 | 25.0 |
| 2015 | 2215 | 424 | 107.6 | 41.8 | 16.1 | 28.0 |
| 2016 | 2967 | 463 | 142.7 | 47.7 | 13.5 | 25.1 |
| 2017 | 2815 | 336 | 130.0 | 31.0 | 10.7 | 19.2 |
| 2018 | 1639 | 291 | 81.1 | 25.9 | 15.1 | 24.2 |

Table 24.2.5. Nephrops in FU22 (Smalls Grounds). Results summary table for geostatistical analysis of UWTV survey.

| Ground | Year | Number of stations | Mean Density adjusted (burrows $/ \mathrm{m}^{2}$ ) | Area Sur- <br> veyed <br> ( $\mathrm{km}^{2}$ ) | Domain <br> Area <br> (km ${ }^{2}$ ) | Burrow <br> Count | Geostatistical Abundance Estimate adjusted (millions burrows) | CV on <br> Burrow estimate \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Smalls | 2006 | 100 | 0.49 | 15 | 2962 | 10,498 | 1503 | 2 |
|  | 2007 | 107 | 0.37 | 16 | 2955 | 8,571 | 1136 | 6 |
|  | 2008 | 76 | 0.36 | 15 | 2698 | 9,411 | 1114 | 6 |
|  | 2009 | 67 | 0.36 | 10 | 2824 | 6,362 | 1093 | 5 |
|  | 2010 | 90 | 0.37 | 15 | 2861 | 8,195 | 1141 | 4 |
|  | 2011 | 107 | 0.41 | 15 | 2881 | 8,191 | 1256 | 3 |
|  | 2012* | 47 | 0.49 | 6 | 2934 | 4,327 | 1498 | 8 |
|  | 2013* | 41 | 0.41 | 7 | 2975 | 3,719 | 1254 | 7 |
|  | 2014* | 52 | 0.53 | 9 | 2970 | 5,715 | 1622 | 8 |
|  | 2015* | 40 | 0.49 | 5 | 3064 | 2,897 | 1363 | 7 |
|  | 2016* | 41 | 0.31 | 6 | 3063 | 2,457 | 866 | 7 |
|  | 2017* | 40 | 0.55 | 6 | 3063 | 4,224 | 1600 | 5 |
|  | 2018* | 42 | 0.31 | 7 | 3063 | 2,534 | 876 | 9 |
|  | 2019* | 41 | 0.40 | 7 | 3063 | 3,573 | 1121 | 6 |

[^13]Table 24.3.1. Nephrops in FU22 (Smalls Grounds). Short-term catch option prediction inputs and recent estimates of mean weight in landings and harvest rate (cells in bold indicates inputs to catch scenrio calculations).

|  |  |  |  |  |  | UWTV abundance estimate |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { 厄 } \\ & \stackrel{y}{\infty} \end{aligned}$ | millions | millions | millions | \% | \% | millions |  | \% | tonnes | tonnes | gramme | gramme |
| 2003 | 95.2 | 67.6 | 145.8 | 34.7 | 41.5 | NA | NA | NA | 2,065 | 720 | 21.7 | 10.7 |
| 2004 | 70.7 | 13.1 | 80.5 | 12.2 | 15.6 | NA | NA | NA | 1,828 | 202 | 25.9 | 15.4 |
| 2005 | 119.3 | 128.6 | 215.7 | 44.7 | 51.9 | NA | NA | NA | 2,533 | 1648 | 21.2 | 12.8 |
| 2006 | 100.2 | 45.2 | 134.1 | 25.3 | 31.1 | 1503 | 70 | 8.9 | 1,761 | 454 | 17.6 | 10.1 |
| 2007 | 165.2 | 180.9 | 300.8 | 45.1 | 52.3 | 1136 | 126 | 26.5 | 2,950 | 1906 | 17.9 | 10.5 |
| 2008 | 143.6 | 26.0 | 163.1 | 12.0 | 15.3 | 1114 | 123 | 14.6 | 3,090 | 289 | 21.5 | 11.1 |
| 2009 | 92.0 | 33.0 | 116.8 | 21.2 | 26.4 | 1093 | 108 | 10.7 | 2,185 | 371 | 23.7 | 11.3 |
| 2010 | 121.8 | 44.5 | 155.2 | 21.5 | 26.8 | 1141 | 88 | 13.6 | 2,714 | 636 | 22.3 | 14.3 |
| 2011 | 60.0 | 13.2 | 69.8 | 14.1 | 18.0 | 1256 | 72 | 5.6 | 1,636 | 196 | 27.3 | 14.9 |
| 2012 | 120.3 | 31.4 | 143.9 | 16.3 | 20.7 | 1498 | 239 | 9.6 | 2,618 | 347 | 21.8 | 11.1 |
| 2013 | 93.5 | 40.1 | 123.6 | 24.3 | 30.0 | 1254 | 177 | 9.9 | 2,257 | 497 | 24.1 | 12.4 |
| 2014 | 100.2 | 33.4 | 125.2 | 20.0 | 25.0 | 1622 | 268 | 7.7 | 2,526 | 460 | 25.2 | 13.8 |


| $\begin{aligned} & \text { 亠丷 } \\ & \stackrel{1}{\infty} \end{aligned}$ |  |  |  |  |  | UWTV abundance estimate |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions | millions | millions | \％ | \％ | millions |  | \％ | tonnes | tonnes | gramme | gramme |
| $\begin{aligned} & \text { 亠ু } \\ & \stackrel{\text { ® }}{2} \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |
|  | millions | millions | millions | \％ | \％ | millions |  | \％ | tonnes | tonnes | gramme | gramme |
| 2015 | 114.1 | 44.4 | 147.4 | 22.6 | 28.0 | 1363 | 180 | 10.8 | 2，350 | 450 | 20.6 | 10.1 |
| 2016 | 160.2 | 53.5 | 200.3 | 20.0 | 25.1 | 866 | 112 | 23.1 | 3，329 | 519 | 20.8 | 9.7 |
| 2017 | 164.4 | 39.2 | 193.7 | 15.2 | 19.2 | 1600 | 153 | 12.1 | 3，560 | 424 | 21.7 | 10.8 |
| 2018 | 97.8 | 31.3 | 121.2 | 19.3 | 24.2 | 876 | 154 | 13.8 | 1，974 | 350 | 20.2 | 11.2 |
| 2019 |  |  |  |  |  | 1121 | 141 |  |  |  |  |  |
|  |  |  | Average 16－18 | 18.2 | 22.8 |  |  |  |  | Average 16－18 | 20.9 | 10.6 |



Figure 24.1.1. Nephrops in FU22 (Smalls Grounds). Time-series of the number of Irish vessels reporting landings of Nephrops from FU22 with a >10 threshold.


Figure 24.1.2. Nephrops in FU22 (Smalls Grounds). Combined box and kite plot of vessel power on the Smalls Grounds by year. The blue line indicates the mean.


Figure 24.2.1. Nephrops in FU22 (Smalls Grounds). Landings in tonnes by country.


Figure 24.2.2. Nephrops in FU22 (Smalls Grounds). Fishing effort Kw days for the Irish otter trawl Nephrops directed fleet ( $\mathbf{3 0 \%}$ of Nephrops weight in total landings).


Figure 24.2.3. Nephrops in FU22 (Smalls Grounds). Sampling levels.


Figure 24.2.4. Nephrops in FU22 (Smalls Grounds). The annual estimated $\mathrm{L}_{50}$ with standard error bounds for the on-board retention ogives for samples from the Smalls grounds. Minimum conservation size (MCR) 25 Carapace Length (mm) shown as dashed line.

## Length frequencies for catch (dotted) and landed(solid):

 Nephrops in FU22

Figure 24.2.5. Nephrops in FU22 (Smalls Grounds). Mean size trends for catches and whole landings by sex over the time-series.


Figure 24.2.6. Nephrops in FU22 (Smalls Grounds). The percentage males in the over the time-series.


Figure 24.2.7. Nephrops in FU22 (Smalls Grounds). Mean weight in catch samples by sex with loess smoother and showing cy clical trends.


Figure 24.2.8. Nephrops in FU22 (Smalls Grounds). Annual mean weights (gr) in the landings and discards.


Figure 24.2.9. Nephrops in FU22 (Smalls Grounds). Contour plots of the krigged density estimates for the UWTV surveys over the time-series.


Figure 24.2.10. Nephrops in FU22 (Smalls Grounds). Time-series of abundance estimates for FU22 (error bars indicate $\mathbf{9 5 \%}$ confidence intervals) and MSY $B_{\text {trigger }}$ is dashed blue line.


## MLS $(25 \mathrm{~mm})$ and 35 mm levels displayed

Figure 24.2.11. Nephrops in FU22 (Smalls Grounds). Mean size trends (Carapace length CLmm) for catches by sex from Irish Groundfish Survey 2003-2018.


Figure 24.3.1. Nephrops in FU22 (Smalls Grounds). Harvest Rate (\% dead removed/UWTV abundance). The dashed and solid lines are the MSY proxy and the harvest rate respectively.

## 22 Plaice in Division 27.7.a (Irish Sea)

## Type of assessment in 2019

WKIrish3 (ICES, 2017) benchmarked this assessment and choose the SAM model, including estimates of discards-at-age into the catch matrix. A baseline run of the model was performed using discards since 1981 reconstructed according to the medium dis-card scenario (ICES, 2017).

## ICES advice applicable to 2018

ICES advises that when the MSY approach is applied, catches in 2018 should be no more than 3336 tonnes. If this stock is not under the EU landing obligation in 2018 and discard rates do not change from the average of the last three years (2014-2016), this implies landings of no more than 1793 tonnes.

## ICES advice applicable to 2019

ICES advises that when the MSY approach is applied, catches in 2019 should be no more than 3503 tonnes.

Last year's advice is available at:
http://ices.dk/sites/pub/Publication\ Reports/Advice/2018/2018/ple.27.7a.pdf

### 22.1 General

## Stock description and management units

The stock assessment area and the management unit are both Division 27.7.a (IrishSea).

## Management applicable in 2018 and 2019

Management of plaice in Division 27.7.a is by TAC and there is a Minimum Conservation Reference Size (MCRS) of 27 cm in force. The agreed TACs and associated implications for plaice in Division 27.7.a are detailed in the tables below.

2018

| Species: | Plaice <br> Pleuronectes platessa |  | Zone: |
| :--- | :--- | :--- | :--- |
| Belgium | 46 | 7a <br> (PLE/07A.) |  |
| France | 20 |  |  |
| Ireland | 1255 |  |  |
| The Netherlands | 14 |  |  |
| United Kingdom | 458 |  |  |
| Union | 1793 |  |  |
| TAC | 1793 |  |  |
|  |  | Analytical TAC |  |
|  |  |  |  |

(Source: Council Regulation (EU) 2018/120, ANNEX IA)

| Species:Plaice <br> Pleuronectes platessa |  | Zone: | 7a <br> (PLE/07A.) |
| :--- | :--- | :--- | :--- |
| Belgium | 134 |  |  |
| France | 58 |  |  |
| Ireland | 1499 |  |  |
| The Netherlands | 41 |  |  |
| United Kingdom | 1343 |  |  |
| Union | 3075 |  |  |
| TAC | 3075 | Analytical TAC |  |
|  |  | Article 7(2) of this Regulation applies |  |

(Source: Council Regulation (EU) 2019/124, ANNEXIA)

## The fishery in 2018

National landings data reported to ICES and Working Group estimates of total landings are given in Table 21.1. A summary by gear is given below.

| Catch (2018) | Landings |  | Discards |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |

The TAC for 2018 was 1793 tonnes and the working group estimate of landings in 2018 was 435 tonnes, which is a $26 \%$ decrease in landings comparable to 2017 and represents $24 \%$ of the 2018 TAC. This shortfall in estimated landings relative to the TAC has occurred in previous years, increasingsteadily from $7 \%$ of the TAC in 2003 to around $70 \%$ in 2008,2009 and 2012 and around $80 \%$ in 2013 and 2014 , before falling to $60 \%$ in 2015 and 2016 . The poor uptake of the quota is not a consequence of an inability to catch sufficient quantities of plaice greater than the MCRS but rather is most likely due to the limited market demand and poor value of the catch.

Landings (based on working group estimates) by the Irish, NI, UK and Belgian fleets comprised approximately $58 \%, 17 \%, 14 \%$ and $10 \%$ respectively of total landings in 2018 . The landings of plaice are mainly split between beam trawlers (44\%; primarily Irish vessels then Belgian vessels) targeting sole, and otter trawlers ( $52 \%$ UK and Irish vessels). Historically, otter trawling was dominated by UK vessels fishing for whitefish, but in recent years, many vessels have switched to target Nephrops (Figure 21.1). Otter trawlers from Ireland and N. Ireland typically target Nephrops in the western Irish Sea.

High levels of discarding are known to occur in all fisheries that catch plaice in the Irish Sea (see Figures 21.4 and 21.5).

A general description of the fishery can be found in the stock annex and also in 'Other Relevant Data' section below. For general mixed fisheries advice applicable to this stock and other species taken in the same fisheries, see Section 6.1.

### 22.2 Data

## Landings

National landings data reported to ICES and Working Group estimates of total landings are given in Table 21.1. The working group procedures used to determine the total international landings numbers- and weights-at-age are documented in the stock annex. As a result of increased rates of discarding, landed numbers-at-age for the younger ages (ages 2 to 4 ) have declined more rapidly over the last two decades than landings of older fish (Figure 21.2a).

## Discards

Discard sampling has been conducted by the UK(E\&W) since 2002 and by Ireland since 1993; Northern Ireland has collected data from 1996 (but not between 2003 and 2005), and Belgium since 2003. Length distributions (LD) of landed and discarded fish estimates for year 2018 are available from InterCatch are presented for UK(E\&W) and Irish otter trawlers targeting demersal fish and to Nephrops (Figure 21.4 and Figure 21.5). For all the fleets except the Irish otter trawlers the pattern is dominated by discarding of small fish, below the MCRS of 27 cm .

WKFLAT (ICES, 2011) first estimated total international discards-at-age and introduced them to the assessment of the stock for the first time. Due to limitations in the data available by gear type, discards for Ireland, France and Northern Ireland, for the years 2004-2011 were raised using UK estimates on the basis of equivalent gear types. A raising factor based on tonnages landed for these countries was calculated and applied to the UK(E\&W) estimates of discard numbers. Finally, these estimates were added to those calculated for Belgium to give estimates of total international discard numbers-at-age.

Since 2012, catch data (landings and discards) are available from InterCatch disaggregated by country and fleet. Total international discards are raised from available discards data.

The total discard estimates (Table 21.1, Figure 21.2b) confirm the significant proportion of discarding that occurs in the fishery, which has increased in time. Since 2004, the majority of the catch has been discarded ( $61 \%$ and $82 \%$ average discard in weight and in numbers respectively, since 2004).

There is a considerablehistoric time period (1981-2003) for which no international raised discard estimates are available. The method for reconstructing discards prior to 2004 is based on sizevarying discard rates and is documented in Annex 4 of the WKIrish3 report (ICES, 2017).

## Biological

Landings numbers-at-age are given in Table 21.5 and plotted in Figure 21.2a. Weights-at-age in the landings are given in Table 21.6. Discard weights-at-age are given in Table 21.7 and weights-at-age in the stock in Table 21.8. The history of the derivation of the landings weights and stock weights used in this assessment is described in the stock annex.

Mean weight-at-age in the landings and survey data indicate declines in both sexes throughout the Irish Sea since 1993 so that plaice at ages $\leq 4$ are typically below MRCS.

## Surveys

All available tuning data are shown in Tables 21.2, 21.3 ( a and b ) and 21.4. Due to inconsistencies in the available commercial tuning fleets, Irish Sea plaice assessments since 2004 have only included the UK (E\&W) beam trawl survey (UK (E\&W)-BTS-Q3) and the two NIGFS-WIBTS spawning biomass indices based on ground fish surveys (NIGFS-WIBTS-Q1 and NIGFS-WIBTSQ4). For more information seeWGNSDS (ICES, 2004). The UK (E\&W)-BTS-Q3 index was revised
by WKFLAT 2011 to include stations in the western Irish Sea and in St George's Channel. A second revision was conducted last year to correct for some inconsistency in the index calculation This revision did not substantially change the trend of the biomass index (see WD Cambiè and Earl, 2017 in WGCSE 2017 report).

Reviews of the UK (E\&W)-BTS-Q3 mean standardised cpue trends have indicated that the survey has good internal consistency in monitoring trends across the stock area. For the entire Irish Sea, the biomass index of ages 1-4 fish calculated from the UK (E\&W)-BTS-Q3 (Figure 21.3, right) indicates two periods of upwards trend, 1993-2003 and from 2007-2015. It is however, detected to have dropped from 2016. An increase of biomass in older ages is observed (Figure 21.3, left). The NIGFS-WIBTS surveys show similar increases in biomass between 1993 and 2003 and then a further increase subsequently until most recent years.
The NIGFS-WIBTS survey strata can be disaggregated into western (Strata 1-3) and eastern (Strata 4-7) subareas, where the subareas are divided by the deep trench that runs roughly northsouth to the west of the Isle of Man (Figure 21.6, Tables 21.3a and b). The opposite trend between spring and autumn estimates of mean biomass in 2018 is mainly driven from eastern (Strata $4-$ 7) subarea.

The SSB of plaice in the Irish Sea is also independently estimated using the Annual Egg Production Method (AEPM), according to Armstrong et al., 2001 methodology.

| Year | SSB (tonnes) | Catch/SSB harvest rate |
| :--- | :---: | :--- | :--- |
| 1995 | 9081 |  |
| 2000 | 13303 | 15.16 |
| 2006 | 14417 | 12.77 |
| 2008 | 14352 | 19.5 |

Catch (discards available from 2004) to egg survey biomass ratios indicate historically that the plaice in the Irish Sea has been lightly exploited. Splitting the SSB estimates from the AEPM into eastern and western Irish Sea areas (Figure 21.7) also indicates that the perceived increase in plaice biomass is due to increased production in the eastern Irish Sea only (for more details see stock annex).

In summary, the UK (E\&W)-BTS-Q3 in September, the NIGFS-WIBTS-Q4 index in October (but not NIGFS-WIBTS-Q1 March), and the AEPM indicate a sustained increase in biomass in the eastern Irish Sea, but this rise does not appear to extend across the deep channel to plaice in the western Irish Sea (Figure 21.7).

## Commercial cpue

Age-based tuning data available for this assessment, comprise three commercial fleets: the UK(E\&W) otter trawl fleet (UK(E\&W) OTB, from 2008), the UK(E\&W) beam trawl fleet (UK(E\&W) BT, from 1989) and the Irish otter trawl fleet (IR-OTB, from 1995). Due to inconsistencies in the available tuning fleets, Irish Sea plaice assessments since 2004 have omitted these indices. For more information, see WGNSDS 2004. The effort and catch by these commercial fleets has been very low in recent years and the cpue data are no longer considered informative.

## Other relevant data

Table 21.2 and Figure 21.1 show that effort levels have decreased since 2002 for the majority of fleets. Both the UK otter and beam trawl fleets are close to their lowest recorded effort levels in time-series extending back to 1972 and 1983 respectively. Effort by UK Nephrops trawlers has greatly increased in the years 2006-2014 but has decreased in the last years. However, this fleet is now the dominant UK fleet in terms of hours fished in 27.7.a. Belgian vessels operating in Division 7 typically move in and out of the Irish Sea, depending on the season, from specifically the Bristol Channel and Celtic Sea, the Bay of Biscay and the southern North Sea.
Since 2013, a problem with the gear effort information (000s hours fished) reported for the UK $(\mathrm{E}+\mathrm{W})$ commercial beam trawl fleet has been registered. Effort information from this fleet is largely missing as a result of a larger component of the fleet using the EU electronic logbook system to report its activities. Gear effort information reporting has not been mandatory with this system to date. As a result, few trips reported their gear effort information rendering the overall effort reported and resultinglpue unusable. However, an initial inspection of an alternate effort indicator for this gear (days fished) suggests that UK beam trawl effort in 2013, 2014, 2015, 2016, and 2017 is at the level observed in 2012. The otter trawl fleet effort reporting was unaffected by this as these vessels were not reporting their landings via this method in these years.

### 22.3 Historical stock development

Model: Age-based analytical assessment (State-space Assessment Model, SAM) that uses landings and discards (Nielsen and Berg, 2014).

Software: R version 3.5 .1 with additional packages (version in parenthesis):
stockassessment (0.8.1); FLCore (2.6.13); reshape (0.8.8); ggplot2 (2.3.1.1); Cairo (1.5-10); doParallel (1.0.14); TMB (1.7.15); devtools (2.0.2).

## Model options chosen

The AP model (Aarts and Poos, 2009) was replaced by SAM. WGCSE (ICES, 2016) agreed that the AP model was not the definitive assessment tool for IrishSea plaice but a temporary solution to the fitting of datasets which included recent discards estimates but for which historic discard information was not available. Reconstructed values of historic discards (prior 2004) were provided in the WKIrish3 (ICES, 2017). TheSAMmodel incorporates the estimated historic discards and is used to run the assessment since 2017.

The model runs were performed using the R package 'stockassessment'. Settings for this update stock assessment are given in the table below. The update assessment follows the same procedure as in the WKIrish3 benchmark assessment. A baseline run of the model was performed using discards since 1981 reconstructed according to the medium discard scenario (ICES, 2017). Discard survival was set at $40 \%$, and natural mortality followed a Lorenzen curve, scaled to 0.12.

## Input data types and characteristics

Commercial catch-at-age data. Discards values available from 2004. Estimates of discards reconstructed for 1981-2003 (WKIrish3). Only the dead fraction of discards (0.6) is accounted for in the model. Three survey indices (UK (E\&W)-BTS-Q3, NIGFS-WIBTS-Q1, and NIGFS-WIBTS-Q4); fixed maturity ogive; natural mortality constant over years and different across ages.

## Final update assessment

WKIrish3 benchmarked this assessment and included estimates of discards-at-age into the catch matrix.

The assessment settings are shown in the following table, with changes to the previous year's settings highlighted in bold. Historic settings are given in the stock annex.

| Assessment year |  | 2015 | 2016 | 2017 | 2018 | 2019 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Assessment model |  | AP | AP | SAM | SAM | SAM |
| Tuning fleets | $\begin{aligned} & \text { UK (E\&W)- } \\ & \text { BTS-Q3 } \end{aligned}$ | Survey omitted | Survey omitted | Survey omitted | Survey omitted | Survey omitted |
|  | Extended UK (E\&W)-BTSQ3 | $\begin{aligned} & \text { 1993-2014, } \\ & \text { ages 1-6 } \end{aligned}$ | $\begin{aligned} & \text { 1993-2015, } \\ & \text { ages } 1-6 \end{aligned}$ | 1993-2016, <br> ages 1-7 | 1993-2017, <br> ages 1-7 | 1993-2018, <br> ages 1-7 |
|  | UK(E\&W) BTS Mar | Survey omitted | Survey omitted | Survey omitted | Survey omitted | Survey omitted |
|  | UK(E\&W) <br> OTB | Series omitted | Series omitted | Series omitted | Series omitted | Series omitted |
|  | UK(E\&W) BT | Series omitted | Series omitted | Series omitted | Series omitted | Series omitted |
|  | IR-OTB | Series omitted | Series omitted | Series omitted | Series omitted | Series omitted |
|  | NIGFS-WI- <br> BTS-Q1 | 1992-2014 | 1992-2015 | 1992-2016 | 1992-2017 | 1992-2018 |
|  | NIGFS-WI- <br> BTS-Q4 | 1992-2014 | 1992-2015 | 1992-2016 | 1992-2017 | 1992-2018 |
| Selectivity model |  | Linear Time Varying Spline at age (TVS) | Linear Time Varying Spline at age (TVS) | Correlated random walk | Correlated random walk | Correlated random walk |
| Discard fraction |  | Polynomial <br> Time Varying <br> Spline at age <br> (PTVS) | Polynomial <br> Time Varying Spline at age (PTVS) | Estimated by WKIRISH3 | Estimated by WKIrish3 | Estimated by WKIRISH3 |
| Landings N at age |  | 1-9+ | 1-9+ | 1-8+ | 1981-2017, <br> ages 1-8+ | 1981-2018, <br> ages 1-8+ |
| Discards N at age |  | $\begin{aligned} & \text { 2004-2013, } \\ & \text { ages 1-5 } \end{aligned}$ | $\begin{aligned} & \text { 2004-2014, } \\ & \text { ages } 1-5 \end{aligned}$ | 1981-2016, <br> ages 1-8+ | $\begin{aligned} & \text { 1981-2017, } \\ & \text { ages 1-8+ } \end{aligned}$ | 1981-2018, <br> ages 1-8+ |

The estimated selectivity patterns split into the landed and discarded components are shown in Figure 21.8. Until early 1990s, the landings selectivity had the highest values for fish aged 4 (indicating that four years age fish were selected). This selectivity shifted to age 5 in late the 1990s and early 2000s, due to the increase of the MCRS in 1998 (from 250 mm to 270 mm ). Since late 2000s landings gradually fell over time to very low values relative to the discard pattern, which became dominant and expanded to the older aged fish during the most recent years.

The catchability of the UK(E\&W)-BTS-Q3 survey is elevated for ages 1 and 2 and reflects the nature of the survey, which was designed as a recruit index (Figure 21.9).

Diagnostic output from the SAMmodel is shown in Figure 21.10. In the catch residuals, negative values are apparent in ages $8+$ from 1998. A year effect in 2004 is present in the UK(E\&W)-BTSQ3 residuals (which is the first year for which discard data are available). A pattern of negative residuals between 2004 and 2009 is present in the residuals of the NIGFS-WIBTS due to large fluctuations in the SSB indices, which are due potentially to variable catchability of the survey.
Recruitment is fluctuating without an overall trend, and it is estimated at its lowest values in 2017 and 2018. The standardised values of the recruitment estimated by the SAM model and the standardised value of age 1 from the UK-BTS survey are characterised by similar pattern, demonstrating consistency in the model estimates (Figure 21.11).

The estimated SSB from the SAM model shows an increasing trend from 1995 until 2004-2005, followed by a drop in 2006 and 2007. This change in SSB trend from 2004 is probably due to the inclusion of more reliable discards values since 2004, when international raised discard estimates became available. Since 2012, SSB has increased reaching the highest value of the whole timeseries in 2016. A slight decrease is observed in estimated SSB in 2017 followed by an increase in 2018. The SSB trend is largely in agreement with independent SSB estimates from the Annual Egg Production Method (AEPM), up to the most recent estimate in 2010, as well as with the survey data used in the assessment (NIGFS-WIBTS-Q1 and -Q4; UK(E\&W)-BTS-Q3, Figure 21.12).

Estimates of numbers-at-age in the landings, discards and population, and fishing mortality numbers-at-age are given in Tables 21.9-21.12. A summary plot for the SAMassessment is shown in Figure 21.13 and the time-series estimates for $\mathrm{Fbar}_{\mathrm{b},} \mathrm{SSB}$ and recruitment are given in Table 21.13.

## Comparison with previous assessments

In 2017, the Aarts and Poos model was replaced by the state-space assessment model (SAM). The assessment used the Lorenzen M scaled to 0.12, and the most recent maturity ogive for the survey.

The methodology provided is as robust as possible and does not currently appear to suffer from a serious retrospective pattern (Figure 21.14). The ten assessment model configurations compared in WKIrish3 perform similarly in terms of temporal trends in SSB, recruitment, catch and $F_{b a r}$ Small retrospective bias in SSB in 2004 likely resulted from the introduction of discards estimates based on samples collected from that year (prior to 2004, discards estimates are reconstructed values based on size-varying discard rates). A Mohn's rho analysis for a five-years peel resulted in values of $0.23 \%$ for recruitment, $4.64 \%$ for SSB and $-5.47 \%$ for $F_{b a r}$.

## State of the stock

Trends in Fbar SSB, recruitment and catch, for the full time-series, are shown in Table 21.13 and Figure 21.13. The assessment consistently estimates that fishing mortality declined from high levels in the 1980s and early 1990s to very low levels, having been $<0.1$ since 2013 . Since 2012, SSB has increased reaching the highest value of the whole time-series in 2016, whereas it has slightly decreased in 2017. Estimated recruitments are highly variable. An increasing trend was
present until 2015 although it seems tohave dropped to the lowest values in 2017 and 2018. Catch has decreased to low levels and, since 2006, the majority of the catch has been discarded ( $61 \%$ in weight and $82 \%$ and number respectively, averaged since 2004).

### 22.4 Short-term projections

Forecasting takes the form of short-term stochastic projections. A total of 1000 samples are generated from the estimated distribution of survivors. These replicates are then simulated forward according to model and forecast assumptions (see table below), using the usual exponential decay equations, but also incorporating the stochastic survival process (using the estimated survival standard deviation) and subject to different catch-options scenarios. Recruitment in the intermediate year (2019) was taken as the median from a distribution about the assessment estimate. Estimates of recruitment for intermediate year and subsequent years were resampled from the 2015-2018 year classes, reflecting recent low levels of recruitment. These re-sampled recruitments are only used for SAM forecasts in order to evaluate future stock dynamics.

| Initial stock size | Starting populations are simulated from the estimated distribution at the start of the inter- <br> mediate year (including covariances) |
| :--- | :--- |
| Maturity | Average of final three years of assessment data |
| Natural mortality | Average of final three years of assessment data |
| F and M before spawn- <br> ing | Both taken as zero |
| Weight at age in the <br> catch | Average of final three years of assessment data |
| Weight at age in the <br> stock | Assumed to be the same as weight-at-age in the catch |
| Exploitation pattern | Fishing mortalities taken as a three-year average <br> Stock recruitment <br> model used <br> Recruitment for the intermediate year onwards is sampled, from 2015 to the final year of <br> catch data. <br> Procedures used for <br> splitting projected <br> catches <br> An average of final three years of landing fractions are used in the forecast period. <br> Discard values are raised to include the live portion. Discard numbers multiplied by 5/3 to ac- <br> count for discard survival. Total catch is sum of three components: landings, discards as- <br> sumed to die, and discards assumed to survive. |

$\overline{\text { F estimates 2016-2018 has fluctuated around similar values, from } 0.050 \text { (2016) to } 0.066 \text { (2017) and }}$ 0.064 (2018). F status quo, Fsq, is set to the same level as F in 2018 and mean $F_{b a r}$ has been estimated by averaging the F over 2016-2018.

A full management options table is provided in Table 21.14. Note that the values that appear in the catch scenarios are medians from the distributions that result from the stochastic forecast. Implementing the management plan for this stock with $\mathrm{F}_{\mathrm{MS}}=0.196$ leads to a total catch of 3299 t ( 1931 t of landings and 1368 t of discards including dead and survivors) in 2020 and SSB of 18354 t in 2021.

### 22.5 Medium-term projections

There are no medium-term projections for this stock.

### 22.6 MSY explorations

The reference points for this stock were estimated in 2018 (ICES, 2018) as ICES request for EU western waters stocks and are presented in the table below.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | $8757$ <br> tonnes | Lower 5th percentile of B $\mathrm{F}_{\text {MSY }}$ | ICES <br> (2018) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.196 | Stochastic simulations with segmented regression from the entire time-series (1981-2017) | ICES <br> (2018) |
| Precautionary approach | $\mathrm{B}_{\text {lim }}$ | $3958$ <br> tonnes | $\mathrm{B}_{\text {loss }}=$ minimum SSB observed | ICES (2018)) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | $5294$ <br> tonnes | $\mathrm{B}_{\text {lim }} \times \exp (1.645 \times \sigma) ; \sigma=0.177$ | $\begin{aligned} & \text { ICES } \\ & \text { (2018) } \end{aligned}$ |
|  | $\mathrm{F}_{\text {lim }}$ | 0. 50 | F with $50 \%$ probability of SSB $<\mathrm{Bl}_{\text {lim }}$ | $\begin{aligned} & \text { ICES } \\ & \text { (2018) } \end{aligned}$ |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.36 | $\mathrm{F}_{\lim } \times \exp (-1.645 \times \sigma) ; \sigma=0.201$ | ICES <br> (2018) |
| Management plan | SSB ${ }_{\text {mgt }}$ | Not applicable |  |  |
|  | $F_{\text {mgt }}$ | Not applicable |  |  |

## Yield per Recruit analysis

There are no yield per recruit analyses for this stock.

### 22.7 Management plans

There are no management plans for this stock.

### 22.8 Uncertainties and bias in assessment and forecast

The assessment was benchmarked in 2017 (WKIrish3), which resulted in the SAM model being fitted using catches based on reconstructed estimates of discards prior to 2004. This discard reconstruction introduces additional uncertainty in the model. The model estimates of stock development since 2004 are more reliable as based on direct discard estimates. The SAM model considered only the dead portion of the discards ( $60 \%$ ), but in the forecast the estimates are raised to include the surviving discards.

### 22.9 Recommendations for next benchmark

There is evidence of substantial substock structure and incorporating information about the differences in growth and maturity between the east and west sides of the Irish Sea, as well as by sex should be explored.

Incorporating data on changes in maturity and natural mortality over time, linked to the decreasing in weights-at-age observed in survey data, should also be considered. There is evidence of a decline in weight-at-age from the commercial landings data and survey data. The UK(E\&W)-BTS-Q3 survey data also indicate declines in length-at-age and maturity-at-age.

Creating age-based indices for the NI groundfish surveys would improve the assessment.
Ecosystem information ought to be explored.

| Year | Candidate <br> Stock | Supporting Justification | Suggested time | Indicate expertise nec- <br> essary at benchmark <br> meeting |
| :--- | :--- | :--- | :--- | :--- |
| 2019 | 27.7.a Plaice | - Incorporating data on changes in maturity and <br> natural mortality over time, linked to the de- <br> creasing in weights-at-age observed in survey <br> data. | when sufficient <br> progress has been <br> made | Expert group members |
|  | - Incorporate information about the differences <br> in growth and maturity be-tween the east and <br> west sides of the Irish Sea, and by sex. |  |  |  |
|  | - Creating age based indices for the NI ground- <br> fish surveys |  |  |  |

### 22.10 Management considerations

The high level of discarding in this fishery indicates a mismatch between the minimum landing size and the mesh size of the gear being used. Any measures that effect a reduction in discards will result in increased future yield. However, the market demand for plaice is poor and small plaice are particularly undesirable. Strong year effects are seen in the discard data and these are likely due to spatial structure in the stock. Spatial management of fleets in the Irish Sea may reduce the discarding of plaice.

The overall state of the stock is consistently estimated to have low fishing mortality and high spawning biomass. Therefore, the stock is considered to be within safe biological limits.

Discarding has increased throughout the period in which data are available, while landings of plaice have decreased, even though the TAC is not restrictive. Effort has decreased in fisheries targeting plaice (including UK(E\&W) and Belgian beam-trawl fisheries, and UK(E\&W) and Irish otter trawl fisheries targeting demersal fish). In contrast, effort by the UK(E\&W) Nephrops fleet has increased, however, this is still small in comparison to effort by the Irish Nephrops fleet. The main Nephrops grounds are located in the western Irish Sea, where relatively small plaice are found. Technical measures to mitigate discarding by all Nephrops fleets could include the use of sorting grids: gear selectivity trials and monitoring from four Irish Nephrops trawlers using grids since 2009 indicate a potential reduction in fish discarding by $75 \%$ (BIM, 2009).

### 22.11 References

Aarts, G., and Poos, J.J. 2009. Comprehensive discard reconstruction and abundance estimation using flexible selectivity functions. ICES Journal of Marine Science, 66: 763-771.

Armstrong M. J., Connolly P., Nash R. D. M., Pawson M. G., Alesworth E., Coulahan P. J., Dickey-Collas M., Milligan S. P., O'Neill M., Witthames P. R. and Woolner L. 2001. An application of the annual egg production method to estimate spawning biomass of cod (Gadusmorhua L.), plaice (Pleuronectes platessa L.) and sole (Solea solea L.) in the Irish Sea. ICES Journal of Marine Science, 58, 183-203.

BIM. 2009. Summary report of Gear Trials to Support Ireland's Submission under Articles $11 \& 13$ of Reg. 1342/2008. Nephrops Fisheries VIIa \& VIIb-k. Project 09.SM. T1.01. Bord Iascaigh Mhara (BIM) May 2009.

Cambiè and Earl, 2017. Review of the UK (E\&W)-BTS-Q3 abundance index for Irish Sea plaice (ple-iris, plaice 7a) Working Document for Working Group on Celtic Seas Ecoregion (WGCSE), 9-18 May, Copenhagen, Denmark. ICES CM 2017/ACOM:13.

ICES. 2004. Report of the Working Group on the assessment of Northern Shelf Demersal Stocks (WGNSDS). ICES Document CM 2005/ACFM: 01. 722 pp.

ICES. 2011. Report of the Benchmark Workshop on Flatfish (WKFLAT), 1-8 February 2011, Copenhagen, Denmark. ICES CM 2011/ACOM:39.

ICES, 2016. Report of the Working Group on Celtic Seas Ecoregion (WGCSE). 4-13 May, Copenhagen, Denmark ICES CM 2016/ACOM:13.

ICES. 2017. Report of the Benchmark Workshop on the Irish Sea Ecosystem (WKIrish3), 30 January-3 February, Galway, Ireland. ICES CM 2017/BSG:01.
ICES. 2018. ICES Special Request Advice Celtic Seas Ecoregion. EU request to ICES to provide plausible and updated FMSY ranges for the stocks of species inhabiting western EU waters. https://doi.org/10.17895/ices.pub. 4149

Nielsen, A., and Berg, C.W. 2014. Estimation of time-varying selectivity in stock assessments using statespace models. Fisheries Research, 158: 96-101.

Table 21.1. Plaice in Division 7.a. History of official landings and ICES estimates of discards. Weights are in tonnes.

| Year |  |  | $\begin{aligned} & \text { ס } \\ & \text { 들 } \\ & \underline{\underline{0}} \end{aligned}$ |  |  |  |  |  | $\begin{aligned} & \text { n } \\ & \frac{0}{0} \\ & \stackrel{y}{0} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 332 | 13 | 547 | - | 1082 | 14 | 63 | 2051 |  |
| 1995 | 327 | 10 | 557 | - | 1050 | 20 | 60 | 2024 |  |
| 1996 | 344 | 11 | 538 | 69 | 878 | 16 | 18 | 1874 |  |
| 1997 | 459 | 8 | 543 | 110 | 798 | 11 | 25 | 1954 |  |
| 1998 | 327 | 8 | 730 | 27 | 679 | 14 | 18 | 1803 |  |
| 1999 | 275 | 5 | 541 | 30 | 687 | 5 | 23 | 1566 |  |
| 2000 | 325 | 14 | 420 | 47 | 610 | 6 | 21 | 1443 |  |
| 2001 | 482 | 9 | 378 | - | 607 | 1 | 11 | 1488 |  |
| 2002 | 636 | 8 | 370 | - | 569 | 1 | 7 | 1591 |  |
| 2003 | 628 | 7 | 490 | - | 409 | 1 | 9 | 1544 |  |
| 2004 | 431 | 2 | 328 | - | 369 | 0 | 4 | 1134 | 1031 |
| 2005 | 566 | 9 | 272 | - | 422 | 0 | 1 | 1270 | 1210 |
| 2006 | 343 | 2 | 179 | 0 | 413 | 0 | 0 | 937 | 1254 |
| 2007 | 194 | 2 | 194 | 0 | 412 | 0 | - | 802 | 1744 |
| 2008 | 157 | 2 | 102 | 0 | 301 | 1 | 1 | 564 | 1268 |
| 2009 | 197 | 0 | 73 | 0 | 187 | 1 | 2 | 460 | 1132 |
| 2010 | 138 | 0 | 89 | 0 | 150 | 0 | 3 | 380 | 2561 |
| 2011 | 332 | 0 | 118 | 0 | 146 | 0 | 0 | 596 | 603 |
| 2012 | 236 | 0 | 108 | 0 | 164 | 0 | 0 | 508 | 1010 |
| 2013 | 144 | 0 | 103 | 0 | 92 | 0 | 0 | 339 | 725 |
| 2014 | 100 | 0 | 123 | 0 | 59 | 0 | 0 | 282 | 943 |
| 2015 | 115 | 0 | 244 | 0 | 80 | 0 | 0 | 439 | 572 |
| 2016 | 82 | 0 | 605 | - | 56 | - | - | 742 | 437 |
| 2017 | 77 | 0 | 446 | - | 62 | - | - | 585* | 852 |
| 2018 | 52 | 0 | 315 | - | 64 | - | - | 432* | 395 |

[^14]Table 21.2. Irish Sea plaice: English standardised lpue and effort, Belgian beam trawl lpue and effort and Irish otter trawl lpue and effort series.

| Year | CPUE |  |  | LPUE |  |  |  |  |  |  | Effort |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | UK(E\&W) Beam trawl survey ${ }^{1}$ <br> March September September <br> Prime only Extended |  |  | UK (E\&W) ${ }^{2}$ |  |  | $\begin{aligned} & \text { Belgian }^{5} \\ & \text { Beam }^{4} \text { Beam } \\ & \text { Trawl Trawl } \end{aligned}$ |  | Irish ${ }^{7}$ |  | UK (E\&W) |  |  |  |  | $\begin{aligned} & \frac{\text { Belgiar }}{\text { Beam }} \frac{\text { Irish }^{9}}{\text { Otter Beam }} \\ & \text { Trawl Trawl Trawl } \end{aligned}$ |  |  |
|  |  |  |  | $\text { Otter }^{3}$ Trawl | $\begin{aligned} & \text { Otter }^{4} \text { Beam } \\ & \\ & \text { Trawl Trawl } \end{aligned}$ |  |  |  | Otter Beam Trawl Trawl |  | $\mathrm{Otter}^{3} \mathrm{Otter}^{4}$ <br> Trawl Trawl |  | Bear Beam ${ }^{4}$ Nephrops $^{3}$ <br> Traw Trawl Trawl |  |  |  |  |  |
| 1972 |  |  |  | 6.96 |  |  |  | 9.8 |  |  | 128.4 |  |  |  |  | 6.8 |  |  |
| 1973 |  |  |  | 6.33 |  |  |  | 9.0 |  |  | 147.6 |  |  |  |  | 16.5 |  |  |
| 1974 |  |  |  | 7.45 |  |  |  | 10.4 |  |  | 115.2 |  |  |  |  | 14.2 |  |  |
| 1975 |  |  |  | 7.71 |  |  |  | 10.7 |  |  | 130.7 |  |  |  |  | 16.2 |  |  |
| 1976 |  |  |  | 5.03 |  |  |  | 5.8 |  |  | 122.3 |  |  |  |  | 15.1 |  |  |
| 1977 |  |  |  | 4.82 |  |  |  | 5.3 |  |  | 101.9 |  |  |  |  | 13.4 |  |  |
| 1978 |  |  |  | 6.77 |  | 4.88 |  | 6.9 |  |  | 89.1 |  | 0.9 |  |  | 12.0 |  |  |
| 1979 |  |  |  | 7.18 |  | 15.23 |  | 8.0 |  |  | 89.9 |  | 1.7 |  |  | 13.7 |  |  |
| 1980 |  |  |  | 8.24 |  | 8.98 |  | 8.6 |  |  | 107.0 |  | 4.3 |  |  | 20.8 |  |  |
| 1981 |  |  |  | 6.87 |  | 4.91 |  | 7.1 |  |  | 107.1 |  | 6.4 |  |  | 26.7 |  |  |
| 1982 |  |  |  | 4.92 |  | 1.77 |  | 4.4 |  |  | 127.2 |  | 5.5 |  |  | 21.3 |  |  |
| 1983 |  |  |  | 5.32 | 1021 | 3.08 | 0 | 7.8 |  |  | 88.1 | 1716.5 | 2.8 | 0 |  | 18.5 |  |  |
| 1984 |  |  |  | 7.77 | 1472 | 6.98 | 810 | 6.8 |  |  | 103.1 | 7932.1 | 4.1 | 263 |  | 13.6 |  |  |
| 1985 |  |  |  | 9.97 | 1946 | 25.70 | 5487 | 8.8 |  |  | 102.9 | 6930.8 | 7.4 | 428.1 |  | 21.9 |  |  |
| 1986 |  |  |  | 9.27 | 1597 | 4.21 | 753 | 8.7 |  |  | 90.3 | 6693.2 | 17.0 | 1122.9 |  | 38.3 |  |  |
| 1987 |  |  |  | 7.20 | 1479 | 3.57 | 963 | 8.2 |  |  | 130.6 | 9008.9 | 22.0 | 1178.5 |  | 43.2 |  |  |
| 1988 |  | 392 |  | 5.02 | 1060 | 3.05 | 743 | 6.3 |  |  | 132.0 | 8292.4 | 18.6 | 1019.2 |  | 32.7 |  |  |
| 1989 |  | 253 |  | 5.51 | 1109 | 13.59 | 2559 | 6.2 |  |  | 139.5 | 16161.4 | 25.3 | 1344.5 |  | 36.7 |  |  |
| 1990 |  | 239 |  | 5.93 | 1074 | 12.02 | 3011 | 7.2 |  |  | 117.1 | 7724.5 | 31.0 | 1473.1 |  | 38.3 |  |  |
| 1991 |  | 157 |  | 4.79 | 916 | 10.56 | 2807 | 7.5 |  |  | 107.3 | 7081.1 | 25.8 | 1211.3 |  | 15.4 |  |  |
| 1992 |  | 188 |  | 4.20 | 719 | 9.99 | 2303 | 11.9 |  |  | 96.8 | 6671.8 | 23.4 | 908.1 |  | 23.0 |  |  |
| 1993 | 91 | 235 | 149 | 3.97 | 667 | 9.50 | 2220 | 5.0 |  |  | 78.9 | 6013.1 | 21.5 | 826.9 |  | 24.4 |  |  |
| 1994 | 128 | 225 | 132 | 4.90 | 770 | 7.79 | 1020 | 9.2 |  |  | 43.0 | 3060 | 20.1 | 1451.6 | 0 | 31.6 |  |  |
| 1995 | 134 | 169 | 109 | 5.08 | 806 | 7.69 | 1001 | 9.5 | 3.2 | 17.3 | 43.1 | 3357 | 20.9 | 1429.4 | 0 | 27.1 | 80.1 | 8.5 |
| 1996 | - 6 | 210 | 111 | 5.37 | 732 | 12.96 | 2587 | 11.8 | 4.1 | 19.0 | 42.2 | 3085.1 | 13.3 | 894.3 | 0 | 22.2 | 64.7 | 6.2 |
| 1997 | 147 | 262 | 148 | 5.25 | 662 | 7.66 | 944 | 13.9 | 3.1 | 13.7 | 39.9 | 2903.3 | 10.8 | 784.4 | 0 | 29.3 | 92.0 | 9.9 |
| 1998 | 113 | 249 | 146 | 5.00 | 657 | 5.66 | 766 | 12.3 | 3.7 | 22.3 | 36.9 | 2620.6 | 10.4 | 696 | 0 | 23.8 | 93.5 | 11.5 |
| 1999 | $-6$ | 264 | 151 | 5.38 | 632 | 7.76 | 895 | 7.1 | 2.3 | 23.2 | 22.9 | 1803.5 | 11.0 | 778.9 | 0 | 37.2 | 109.7 | 14.7 |
| 2000 | $-{ }^{6}$ | 357 | 169 | 5.02 | 828 | 13.04 | 1773 | 7.8 | 2.0 | 13.8 | 27.0 | 2034.9 | 6.3 | 410.7 | 0 | 27.0 | 82.6 | 11.4 |
| 2001 |  | 281 | 147 | 3.35 | 539 | 8.33 | 1017 | 9.2 | 2.9 | 14.0 | 33.0 | 2352.9 | 12.5 | 767.4 | 0 | 41.9 | 77.4 | 13.1 |
| 2002 |  | 340 | 200 | 5.66 | 840 | 5.46 | 445 | 7.4 | 2.8 | 7.9 | 24.8 | 1774 | 8.0 | 535.1 | 0 | 52.5 | 77.4 | 17.7 |
| 2003 |  | 503 | 247 | 2.60 | 414 | 3.76 | 400 | 7.5 | 4.1 | 9.5 | 23.9 | 1728.3 | 14.0 | 863.7 | 0 | 48.7 | 73.8 | 18.6 |
| 2004 |  | 540 | 249 | 3.17 | 472 | 4.20 | 255 | 11.2 | 2.1 | 8.6 | 23.5 | 1727 | 7.4 | 419.9 | 0 | 36.1 | 72.5 | 14.2 |
| 2005 |  | 367 | 177 | 4.85 | 540 | 4.67 | 381 | 12.8 | 2.0 | 8.0 | 16.7 | 1313.6 | 11.6 | 627.8 | 1 | 42.1 | 69 | 14.7 |
| 2006 |  | 356 | 166 | 6.50 | 610 | 2.19 | 202 | 10.8 | 1.4 | 6.2 | 5.2 | 478.5 | 4.6 | 280.1 | 10.9 | 28.9 | 66.8 | 12.2 |
| 2007 |  | 432 | 190 | 17.94 | 756 | 4.22 | 550 | 6.9 | 1.3 | 6.1 | 4.4 | 397.2 | 3.2 | 193.5 | 12.6 | 23.8 | 75.9 | 14.2 |
| 2008 |  | 416 | 189 | 9.03 | 469 | 4.47 | 267 | 9.5 | 0.9 | 5.1 | 2.7 | 320.4 | 1.3 | 98 | 11.5 | 12.4 | 59.9 | 9.5 |
| 2009 |  | 467 | 199 | 6.46 | 338 | 1.21 | 169 | 10.1 | 1.1 | 3.8 | 1.5 | 157.7 | 0.46 | 24.9 | 10.0 | 14.7 | 42.8 | 7.6 |
| 2010 |  | 400 | 164 | 11.55 | 371 | 14.39 | 151 | 7.9 | 1.0 | 4.8 | 1.0 | 151 | 0.19 | 10.2 | 9.2 | 15.2 | 45.8 | 9.4 |
| 2011 |  | 417 | 140 | 4.35 | 183 | 11.95 | 701 | 17.3 | 1.2 | 6.8 | 0.69 | 72.7 | 1.56 | 91.2 | 8.6 | 16.4 | 54.5 | 8.1 |
| 2012 |  | 460 | 188 | 0.74 | 276 | 7.25 | 164 | 14.9 | 1.0 | 5.0 | 0.4 | 85 | 0.9 | 60.7 | 12.1 | 14.5 | 58.3 | 7.2 |
| 2013 |  | 550 | 207 | 7.41 | 236 | -8 | 0 | 14.0 | 1.6 | 5.4 | 0.3 | 31.9 | -8 | 1.3 | 10.6 | 8.9 | 42.6 | 5.0 |
| 2014 |  | 592 | 255 | - | 87 | -8 | 0 | 13.9 | 1.5 | 8.3 | - | 16.1 | -8 | 0.4 | 8.3 | 5.1 | 47.8 | 6.0 |
| 2015 |  | 564 | 230 | - | 0 | -8 | 48 | 20.4 | 3.3 | 8.6 | - | 0 | -8 | 0.9 | 4.5 | 4.6 | 39.8 | 8.3 |
| 2016 |  | 582 | 220 | - | 0 | -8 | 0 | 26.4 | 4.6 | 32.8 | - | 0 | -8 | 3.9 | 2.5 | 2.5 | 33.4 | 7.9 |
| 2017 |  | 525 | 170 | - | 244 | -8 | 0 | 17.1 | 11.3 | 35.4 | - | 160.7 | -8 | 0 | 0.3 | 4.2 | 12.1 | 7.5 |
| 2018 |  | 554 | 139 | - | 237 | -8 | 0 | - | 8.4 | 19.5 | - | 238 | -8 | 0 | - | - | 13.6 | 9.6 |

$1 \mathrm{Kg} / 100 \mathrm{~km}$. Sept Prime: ISS/ISN Traditional Prime Stations Only. Sept Extended: ISS/ISN/ISW/SGC All Stations.
2 Whole weight (kg) per corrected hour fished, weighted by area
3 '000 hours fished (corrected for fishing power GRT)
4 days fished
5 Corrected for fishing power (HP) [data for 1999-2010, replaced at 2011WG following recalculation at WKFLAT 2011]
6 Carhelmar survey, $\mathrm{Kg} / 100 \mathrm{~km}$ not available
7 All years updated in 2007 due to slight historical differences
8 Effort not reported in hours for this fleet, see Section 6.7.2 for more detail
9 '000s hours
Fishing power corrections are detailed in Appendix 2 of the 2000 working group report

Table 21.3a. Irish Sea plaice: NIGFS-WIBTS-Q1 indices of relative biomass trends by region in spring.

| NIGFS-WIBTS-Q1 <br> Mar (Spring) | ESTIMATED MEAN ABUNDANCE (kg/3 miles) |  |  | ESTIMATED STANDARD ERROR |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Combined | West | East | Combined | West | East |
| Year | Str 1-7 | Str 1-3 | Str 4-7 | Str 1-7 | Str 1-3 | Str 4-7 |
| 1992 | 8.35 | 5.47 | 9.20 | 3.45 | 1.96 | 4.44 |
| 1993 | 12.36 | 18.43 | 10.54 | 2.14 | 4.78 | 2.39 |
| 1994 | 9.65 | 4.47 | 11.09 | 2.43 | 1.46 | 3.12 |
| 1995 | 7.27 | 4.79 | 7.64 | 1.24 | 0.83 | 1.59 |
| 1996 | 7.29 | 12.60 | 5.70 | 1.64 | 5.71 | 1.28 |
| 1997 | 13.87 | 14.72 | 13.54 | 3.19 | 5.68 | 3.77 |
| 1998 | 10.40 | 13.32 | 9.00 | 2.73 | 7.10 | 2.84 |
| 1999 | 10.71 | 13.53 | 9.59 | 1.81 | 4.92 | 1.84 |
| 2000 | 12.92 | 26.29 | 8.88 | 4.11 | 17.00 | 1.66 |
| 2001 | 12.06 | 18.03 | 9.92 | 1.41 | 4.25 | 1.31 |
| 2002 | 15.27 | 27.95 | 11.17 | 2.53 | 8.39 | 2.14 |
| 2003 | 20.97 | 40.71 | 15.09 | 6.11 | 23.98 | 3.44 |
| 2004 | 8.55 | 5.69 | 9.40 | 1.74 | 1.21 | 2.24 |
| 2005 | 11.10 | 19.43 | 8.62 | 1.93 | 5.99 | 1.76 |
| 2006 | 7.85 | 12.14 | 6.39 | 1.39 | 4.62 | 1.16 |
| 2007 | 6.25 | 14.47 | 3.80 | 1.27 | 4.80 | 0.83 |
| 2008 | 4.46 | 5.11 | 4.57 | 0.76 | 1.23 | 0.91 |
| 2009 | 7.90 | 7.85 | 7.86 | 1.27 | 2.04 | 1.53 |
| 2010 | 19.40 | 8.77 | 17.30 | 1.86 | 2.70 | 2.28 |
| 2011 | 16.34 | 26.20 | 13.03 | 3.51 | 10.11 | 3.41 |
| 2012 | 14.22 | 21.47 | 11.05 | 2.37 | 7.48 | 2.13 |
| 2013 | 21.89 | 28.98 | 16.57 | 3.74 | 8.04 | 4.21 |
| 2014 | 11.43 | 10.96 | 9.65 | 2.04 | 4.82 | 2.22 |
| 2015 | 22.81 | 22.57 | 18.66 | 2.84 | 7.18 | 3.01 |
| 2016 | 34.52 | 30.29 | 35.77 | 7.17 | 9.95 | 8.82 |
| 2017 | 16.10 | 14.85 | 16.47 | 3.16 | 3.90 | 3.70 |
| 2018 | 19.26 | 22.86 | 18.18 | 4.11 | 10.19 | 4.39 |
| 2019 | 5.47 | 6.61 | 5.14 | 1.14 | 2.06 | 1.34 |

Table 21.3b. Irish Sea plaice: NIGFS-WIBTS-Q4 indices of relative biomass trends by region in autumn.

| NIGFS-WIBTS-Q4 <br> Oct (Autumn) | ESTIMATED MEAN ABUNDANCE (kg/3 miles) |  |  | ESTIMATED STANDARD ERROR |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Combined | West | East | Combined | West | East |
| Year | Str 1-7 | Str 1-3 | Str 4-7 | Str 1-7 | Str 1-3 | Str 4-7 |
| 1992 | 4.81 | 2.31 | 5.55 | 0.92 | 1.10 | 1.15 |
| 1993 | 4.48 | 2.08 | 5.20 | 1.00 | 0.87 | 1.27 |
| 1994 | 8.73 | 5.49 | 9.69 | 2.30 | 2.83 | 2.86 |
| 1995 | 4.17 | 5.50 | 3.77 | 1.13 | 2.23 | 1.31 |
| 1996 | 8.68 | 8.85 | 8.63 | 2.25 | 5.94 | 2.33 |
| 1997 | 7.93 | 5.76 | 8.58 | 2.24 | 2.59 | 2.80 |
| 1998 | 5.33 | 3.68 | 5.82 | 1.46 | 2.48 | 1.74 |
| 1999 | 5.81 | 4.30 | 6.26 | 1.67 | 3.08 | 1.97 |
| 2000 | 9.75 | 2.20 | 12.00 | 5.76 | 1.13 | 7.47 |
| 2001 | 13.85 | 2.30 | 17.30 | 6.57 | 1.67 | 8.51 |
| 2002 | 9.80 | 5.90 | 10.97 | 3.91 | 3.61 | 4.97 |
| 2003 | 18.01 | 7.52 | 21.14 | 5.84 | 4.16 | 7.48 |
| 2004 | 7.79 | 1.64 | 9.63 | 1.80 | 0.81 | 2.33 |
| 2005 | 11.35 | 3.41 | 13.72 | 4.51 | 2.18 | 5.82 |
| 2006 | 6.61 | 2.56 | 7.82 | 1.53 | 1.42 | 1.94 |
| 2007 | 7.15 | 4.07 | 8.07 | 1.41 | 2.00 | 1.73 |
| 2008 | 8.68 | 3.28 | 10.27 | 2.20 | 2.09 | 2.78 |
| 2009 | 12.44 | 4.06 | 15.01 | 2.59 | 3.12 | 3.23 |
| 2010 | 15.58 | 5.83 | 18.53 | 5.26 | 5.21 | 6.65 |
| 2011 | 14.48 | 5.39 | 15.94 | 3.55 | 2.66 | 4.55 |
| 2012 | 16.05 | 17.89 | 15.65 | 4.43 | 11.16 | 4.68 |
| 2013 | 17.90 | 13.55 | 19.09 | 4.33 | 11.27 | 4.51 |
| 2014 | 22.18 | 27.67 | 20.35 | 7.61 | 24.88 | 6.52 |
| 2015 | 18.21 | 11.15 | 20.31 | 4.39 | 8.76 | 5.06 |
| 2016 | 17.57 | 0.95 | 22.53 | 4.52 | 0.43 | 5.86 |
| 2017 | 18.55 | 2.96 | 23.20 | 4.25 | 1.59 | 5.50 |
| 2018 | 7.21 | 6.89 | 7.30 | 1.86 | 6.08 | 1.59 |

Table 21.4. Irish Sea plaice: UK (E\&W)-BTS-Q3 biomass index (extended area). Ages in bold are those used in the assessment (ages 1-7).

| Year | Distance towed (kms) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 292.77 | 0.13 | 4.64 | 4.03 | 0.82 | 0.43 | 0.03 | 0.04 | 0.08 | 0.01 | 0.02 |
| 1994 | 218.65 | 0.33 | 4.13 | 2.48 | 1.42 | 0.28 | 0.10 | 0.03 | 0.02 | 0.03 | 0.04 |
| 1995 | 218.65 | 0.78 | 5.56 | 1.96 | 0.84 | 0.41 | 0.07 | 0.05 | 0.02 | 0.00 | 0.03 |
| 1996 | 222.36 | 0.26 | 5.79 | 2.17 | 0.53 | 0.19 | 0.20 | 0.05 | 0.02 | 0.00 | 0.02 |
| 1997 | 218.65 | 0.96 | 5.47 | 2.91 | 1.26 | 0.30 | 0.16 | 0.17 | 0.05 | 0.02 | 0.03 |
| 1998 | 218.65 | 0.56 | 4.50 | 4.26 | 1.09 | 0.38 | 0.21 | 0.08 | 0.06 | 0.01 | 0.04 |
| 1999 | 214.95 | 1.86 | 3.96 | 3.91 | 1.99 | 0.68 | 0.29 | 0.09 | 0.07 | 0.03 | 0.05 |
| 2000 | 218.65 | 1.22 | 8.74 | 2.80 | 1.47 | 1.11 | 0.47 | 0.12 | 0.09 | 0.03 | 0.04 |
| 2001 | 214.95 | 0.83 | 5.99 | 3.62 | 1.11 | 0.60 | 0.54 | 0.11 | 0.06 | 0.02 | 0.01 |
| 2002 | 214.95 | 0.23 | 6.46 | 4.94 | 2.27 | 0.88 | 0.53 | 0.48 | 0.10 | 0.04 | 0.04 |
| 2003 | 211.24 | 2.07 | 6.12 | 5.85 | 2.61 | 1.58 | 0.58 | 0.38 | 0.25 | 0.07 | 0.07 |
| 2004 | 214.95 | 1.09 | 8.07 | 5.36 | 3.94 | 1.88 | 1.15 | 0.21 | 0.19 | 0.13 | 0.10 |
| 2005 | 211.24 | 1.75 | 3.76 | 4.75 | 1.98 | 1.42 | 0.80 | 0.48 | 0.11 | 0.09 | 0.06 |
| 2006 | 214.95 | 3.56 | 5.01 | 3.45 | 2.46 | 1.10 | 0.79 | 0.36 | 0.20 | 0.02 | 0.07 |
| 2007 | 214.95 | 1.15 | 7.97 | 4.47 | 1.66 | 1.20 | 0.65 | 0.33 | 0.25 | 0.14 | 0.06 |
| 2008 | 200.12 | 1.22 | 4.68 | 5.71 | 2.03 | 1.15 | 0.82 | 0.31 | 0.12 | 0.08 | 0.05 |
| 2009 | 214.95 | 1.23 | 4.74 | 3.40 | 3.30 | 0.99 | 0.66 | 0.63 | 0.16 | 0.11 | 0.20 |
| 2010 | 211.24 | 2.01 | 6.22 | 4.31 | 2.05 | 1.44 | 0.66 | 0.54 | 0.36 | 0.20 | 0.19 |
| 2011 | 211.24 | 1.02 | 6.73 | 4.28 | 1.75 | 1.00 | 1.08 | 0.47 | 0.27 | 0.24 | 0.37 |
| 2012 | 214.95 | 1.40 | 6.52 | 6.37 | 1.71 | 1.03 | 0.47 | 0.53 | 0.30 | 0.14 | 0.42 |
| 2013 | 214.95 | 2.04 | 4.33 | 5.05 | 3.08 | 1.60 | 1.07 | 0.47 | 0.44 | 0.20 | 0.42 |
| 2014 | 214.95 | 1.56 | 7.82 | 6.85 | 3.13 | 2.16 | 0.99 | 0.77 | 0.44 | 0.20 | 0.28 |
| 2015 | 214.95 | 1.02 | 6.16 | 6.88 | 2.60 | 1.80 | 1.04 | 0.66 | 0.37 | 0.19 | 0.50 |
| 2016 | 211.24 | 0.18 | 2.91 | 5.97 | 3.95 | 2.45 | 1.61 | 0.96 | 0.74 | 0.45 | 0.58 |
| 2017 | 214.95 | 0.03 | 1.35 | 4.77 | 2.81 | 2.23 | 1.84 | 0.75 | 0.59 | 0.38 | 0.26 |
| 2018 | 214.95 | 0.36 | 1.97 | 2.75 | 2.28 | 1.51 | 1.37 | 1.24 | 0.75 | 0.56 | 0.27 |

Table 21.5. Irish Sea plaice: Landings number-at-age 1 to 8+ (thousands), where rows are years 1981-2018 and columns are ages 1 to $8+$.

| IRISH SEA PLAICE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 12 |  |  |  |  |  |  |  |
| 19812018 |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| 22 | 1742 | 5939 | 2984 | 837 | 222 | 105 | 236 |
| 27 | 715 | 3288 | 3082 | 1358 | 330 | 137 | 213 |
| 51 | 2924 | 2494 | 3211 | 1521 | 648 | 211 | 252 |
| 41 | 3159 | 5179 | 1182 | 1054 | 459 | 299 | 252 |
| 4 | 2357 | 6152 | 3301 | 614 | 429 | 262 | 340 |
| 31 | 1652 | 5280 | 2942 | 1287 | 344 | 371 | 308 |
| 62 | 3717 | 5317 | 5252 | 1341 | 1072 | 123 | 338 |
| 46 | 2923 | 5040 | 2552 | 1400 | 750 | 316 | 405 |
| 24 | 1735 | 5945 | 2671 | 854 | 436 | 214 | 364 |
| 15 | 1019 | 2715 | 2935 | 1132 | 465 | 259 | 223 |
| 180 | 2008 | 1506 | 1929 | 1205 | 465 | 182 | 226 |
| 151 | 1958 | 3209 | 1435 | 1358 | 903 | 388 | 294 |
| 28 | 910 | 1649 | 1357 | 474 | 556 | 377 | 302 |
| 97 | 1146 | 2173 | 1309 | 644 | 318 | 245 | 263 |
| 21 | 961 | 1703 | 1936 | 764 | 318 | 138 | 157 |
| 37 | 856 | 1345 | 1196 | 943 | 370 | 128 | 135 |
| 28 | 830 | 1590 | 1513 | 1003 | 482 | 285 | 257 |
| 6 | 691 | 1739 | 1025 | 612 | 476 | 403 | 385 |
| 68 | 803 | 1505 | 1294 | 696 | 280 | 196 | 242 |
| 0 | 450 | 1174 | 1284 | 686 | 212 | 219 | 203 |
| 14 | 374 | 1138 | 1083 | 767 | 409 | 179 | 166 |
| 1 | 206 | 940 | 1482 | 842 | 539 | 318 | 170 |
| 0 | 286 | 1031 | 1314 | 707 | 415 | 253 | 222 |
| 8 | 198 | 967 | 1104 | 705 | 247 | 114 | 186 |
| 6 | 228 | 708 | 1177 | 890 | 461 | 204 | 213 |
| 5 | 180 | 620 | 550 | 684 | 346 | 220 | 218 |
| 0 | 64 | 351 | 860 | 507 | 401 | 151 | 164 |
| 1 | 99 | 386 | 389 | 409 | 215 | 141 | 119 |
| 0 | 13 | 204 | 374 | 351 | 272 | 117 | 120 |
| 0 | 7 | 75 | 271 | 306 | 193 | 160 | 115 |
| 2 | 53 | 199 | 357 | 483 | 305 | 194 | 191 |
| 0 | 8 | 150 | 292 | 301 | 367 | 218 | 226 |
| 1 | 16 | 87 | 203 | 166 | 149 | 144 | 165 |
| 3 | 6 | 65 | 165 | 160 | 143 | 70 | 158 |
| 0 | 1 | 43 | 93 | 185 | 210 | 149 | 349 |
| 14 | 14 | 58 | 162 | 224 | 346 | 180 | 482 |
| 5 | 4 | 24 | 145 | 206 | 241 | 209 | 520 |
| 0 | 6 | 84 | 110 | 201 | 178 | 151 | 358 |

Table 21.6. Irish Sea plaice: Landings weight-at-age 1 to $8+(\mathrm{kg})$, where rows are years 1981-2018 and columns are ages 1 to $8+$

| IRISH SEA PLAICE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 13 |  |  |  |  |  |  |  |
| 19812018 |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| 0.069 | 0.176 | 0.267 | 0.376 | 0.512 | 0.592 | 0.678 | 1.085 |
| 0.201 | 0.274 | 0.284 | 0.348 | 0.421 | 0.545 | 0.650 | 0.889 |
| 0.232 | 0.261 | 0.290 | 0.319 | 0.368 | 0.426 | 0.484 | 0.699 |
| 0.260 | 0.290 | 0.330 | 0.380 | 0.470 | 0.560 | 0.660 | 0.964 |
| 0.290 | 0.310 | 0.340 | 0.390 | 0.470 | 0.540 | 0.630 | 0.851 |
| 0.270 | 0.280 | 0.340 | 0.420 | 0.500 | 0.540 | 0.630 | 0.980 |
| 0.260 | 0.290 | 0.315 | 0.370 | 0.440 | 0.520 | 0.610 | 0.916 |
| 0.230 | 0.260 | 0.300 | 0.370 | 0.460 | 0.550 | 0.680 | 1.243 |
| 0.227 | 0.272 | 0.321 | 0.374 | 0.430 | 0.491 | 0.555 | 0.761 |
| 0.200 | 0.257 | 0.316 | 0.376 | 0.439 | 0.504 | 0.570 | 0.747 |
| 0.247 | 0.267 | 0.295 | 0.332 | 0.377 | 0.431 | 0.494 | 0.652 |
| 0.169 | 0.218 | 0.274 | 0.337 | 0.407 | 0.484 | 0.568 | 0.799 |
| 0.260 | 0.270 | 0.292 | 0.328 | 0.375 | 0.436 | 0.508 | 0.690 |
| 0.156 | 0.207 | 0.268 | 0.338 | 0.416 | 0.504 | 0.600 | 0.816 |
| 0.189 | 0.224 | 0.262 | 0.329 | 0.353 | 0.406 | 0.461 | 0.699 |
| 0.204 | 0.223 | 0.270 | 0.333 | 0.398 | 0.493 | 0.584 | 0.837 |
| 0.205 | 0.233 | 0.241 | 0.286 | 0.354 | 0.410 | 0.510 | 0.620 |
| 0.185 | 0.226 | 0.249 | 0.316 | 0.353 | 0.410 | 0.468 | 0.655 |
| 0.205 | 0.236 | 0.250 | 0.300 | 0.375 | 0.457 | 0.483 | 0.615 |
| 0.000 | 0.259 | 0.270 | 0.307 | 0.337 | 0.429 | 0.437 | 0.623 |
| 0.232 | 0.233 | 0.271 | 0.334 | 0.396 | 0.439 | 0.571 | 0.764 |
| 0.228 | 0.271 | 0.267 | 0.308 | 0.386 | 0.476 | 0.518 | 0.673 |
| 0.000 | 0.235 | 0.289 | 0.335 | 0.383 | 0.458 | 0.567 | 0.678 |
| 0.214 | 0.239 | 0.258 | 0.297 | 0.347 | 0.416 | 0.543 | 0.571 |
| 0.235 | 0.245 | 0.265 | 0.292 | 0.322 | 0.394 | 0.441 | 0.632 |
| 0.200 | 0.256 | 0.265 | 0.282 | 0.321 | 0.378 | 0.425 | 0.568 |
| 0.000 | 0.280 | 0.266 | 0.281 | 0.320 | 0.371 | 0.416 | 0.481 |
| 0.246 | 0.228 | 0.257 | 0.281 | 0.311 | 0.364 | 0.431 | 0.553 |
| 0.000 | 0.257 | 0.256 | 0.265 | 0.305 | 0.330 | 0.395 | 0.482 |
| 0.000 | 0.260 | 0.265 | 0.282 | 0.301 | 0.356 | 0.392 | 0.492 |
| 0.236 | 0.251 | 0.257 | 0.283 | 0.298 | 0.354 | 0.404 | 0.513 |
| 0.117 | 0.259 | 0.254 | 0.281 | 0.299 | 0.318 | 0.345 | 0.430 |
| 0.249 | 0.245 | 0.249 | 0.267 | 0.297 | 0.330 | 0.386 | 0.417 |
| 0.181 | 0.250 | 0.282 | 0.300 | 0.336 | 0.373 | 0.457 | 0.492 |
| NA | 0.183 | 0.264 | 0.287 | 0.299 | 0.340 | 0.403 | 0.617 |
| 0.113 | 0.149 | 0.229 | 0.318 | 0.422 | 0.362 | 0.433 | 0.660 |
| 0.166 | 0.222 | 0.273 | 0.345 | 0.370 | 0.405 | 0.442 | 0.505 |
| 0.000 | 0.292 | 0.327 | 0.353 | 0.345 | 0.398 | 0.399 | 0.465 |

Table 21.7. Irish Sea plaice: Discards weight-at-age 1 to $8+(\mathrm{kg})$, where rows are years 1981-2018 and columns are ages 1 to $8+$.

| IRISH SEA PLAICE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 123 |  |  |  |  |  |  |  |
| 19812018 |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.057 | 0.115 | 0.145 | 0.164 | 0.211 | 0.290 | 0.238 | 0.210 |
| 0.099 | 0.117 | 0.134 | 0.179 | 0.178 | 0.277 | 0.644 | 0.356 |
| 0.141 | 0.113 | 0.141 | 0.145 | 0.162 | 0.210 | 0.274 | 0.077 |
| 0.044 | 0.081 | 0.113 | 0.140 | 0.150 | 0.205 | 0.219 | 0.243 |
| 0.096 | 0.097 | 0.116 | 0.135 | 0.151 | 0.173 | 0.217 | 0.170 |
| 0.033 | 0.080 | 0.119 | 0.147 | 0.165 | 0.196 | 0.232 | 0.276 |
| 0.083 | 0.101 | 0.138 | 0.183 | 0.201 | 0.140 | 0.194 | 0.225 |
| 0.077 | 0.098 | 0.116 | 0.141 | 0.157 | 0.168 | 0.164 | 0.176 |
| 0.026 | 0.038 | 0.081 | 0.119 | 0.162 | 0.200 | 0.157 | 0.182 |
| 0.064 | 0.069 | 0.094 | 0.116 | 0.144 | 0.157 | 0.181 | 0.181 |
| 0.056 | 0.067 | 0.084 | 0.120 | 0.128 | 0.150 | 0.152 | 0.153 |
| 0.088 | 0.059 | 0.079 | 0.101 | 0.095 | 0.126 | 0.152 | 0.136 |
| 0.136 | 0.103 | 0.109 | 0.120 | 0.146 | 0.161 | 0.155 | 0.170 |
| 0.093 | 0.080 | 0.118 | 0.124 | 0.128 | 0.153 | 0.137 | 0.157 |
| 0.022 | 0.053 | 0.075 | 0.109 | 0.142 | 0.143 | 0.146 | 0.202 |

Table 21.8. Irish Sea plaice: New stock weights-at-age modified to include discard element (kg), where rows are years 1981-2018 and columns are ages 1 to 8+.

| IRISH SEA PLAICE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 14 |  |  |  |  |  |  |  |
| 19812018 |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| 0.087 | 0.124 | 0.190 | 0.351 | 0.509 | 0.592 | 0.678 | 1.085 |
| 0.091 | 0.141 | 0.210 | 0.327 | 0.418 | 0.545 | 0.650 | 0.889 |
| 0.097 | 0.173 | 0.231 | 0.303 | 0.366 | 0.426 | 0.484 | 0.699 |
| 0.100 | 0.196 | 0.275 | 0.362 | 0.467 | 0.560 | 0.660 | 0.964 |
| 0.089 | 0.203 | 0.293 | 0.374 | 0.468 | 0.540 | 0.630 | 0.851 |
| 0.098 | 0.171 | 0.292 | 0.401 | 0.497 | 0.540 | 0.630 | 0.980 |
| 0.102 | 0.208 | 0.266 | 0.353 | 0.437 | 0.519 | 0.610 | 0.916 |
| 0.104 | 0.171 | 0.250 | 0.351 | 0.456 | 0.549 | 0.680 | 1.243 |
| 0.100 | 0.183 | 0.261 | 0.352 | 0.425 | 0.490 | 0.555 | 0.761 |
| 0.090 | 0.172 | 0.253 | 0.349 | 0.431 | 0.502 | 0.570 | 0.747 |
| 0.140 | 0.165 | 0.230 | 0.305 | 0.369 | 0.429 | 0.494 | 0.652 |
| 0.106 | 0.159 | 0.209 | 0.302 | 0.395 | 0.481 | 0.568 | 0.799 |
| 0.097 | 0.141 | 0.209 | 0.291 | 0.363 | 0.434 | 0.508 | 0.690 |
| 0.101 | 0.134 | 0.193 | 0.299 | 0.400 | 0.501 | 0.600 | 0.816 |
| 0.091 | 0.138 | 0.184 | 0.289 | 0.340 | 0.404 | 0.461 | 0.699 |
| 0.091 | 0.130 | 0.181 | 0.286 | 0.377 | 0.488 | 0.583 | 0.837 |
| 0.091 | 0.118 | 0.168 | 0.247 | 0.335 | 0.406 | 0.509 | 0.620 |
| 0.088 | 0.116 | 0.148 | 0.223 | 0.305 | 0.399 | 0.466 | 0.655 |
| 0.100 | 0.125 | 0.150 | 0.216 | 0.321 | 0.444 | 0.480 | 0.615 |
| NA | 0.121 | 0.157 | 0.222 | 0.300 | 0.420 | 0.436 | 0.623 |
| 0.091 | 0.119 | 0.161 | 0.239 | 0.352 | 0.431 | 0.569 | 0.764 |
| 0.088 | 0.114 | 0.161 | 0.228 | 0.347 | 0.467 | 0.517 | 0.673 |
| NA | 0.115 | 0.165 | 0.234 | 0.335 | 0.448 | 0.566 | 0.678 |
| 0.070 | 0.131 | 0.169 | 0.217 | 0.304 | 0.407 | 0.540 | 0.570 |
| 0.103 | 0.127 | 0.161 | 0.238 | 0.234 | 0.377 | 0.454 | 0.602 |
| 0.141 | 0.122 | 0.162 | 0.175 | 0.256 | 0.323 | 0.417 | 0.564 |
| 0.044 | 0.084 | 0.123 | 0.167 | 0.209 | 0.290 | 0.335 | 0.377 |
| 0.096 | 0.100 | 0.131 | 0.168 | 0.204 | 0.279 | 0.397 | 0.285 |
| 0.033 | 0.081 | 0.125 | 0.173 | 0.213 | 0.266 | 0.333 | 0.413 |
| 0.083 | 0.101 | 0.140 | 0.191 | 0.211 | 0.190 | 0.226 | 0.290 |
| 0.078 | 0.104 | 0.137 | 0.182 | 0.221 | 0.271 | 0.334 | 0.364 |
| 0.026 | 0.038 | 0.088 | 0.142 | 0.199 | 0.246 | 0.232 | 0.294 |
| 0.065 | 0.071 | 0.098 | 0.133 | 0.185 | 0.240 | 0.292 | 0.363 |
| 0.056 | 0.068 | 0.089 | 0.135 | 0.153 | 0.194 | 0.214 | 0.296 |
| 0.088 | 0.060 | 0.083 | 0.115 | 0.130 | 0.163 | 0.269 | 0.515 |
| 0.133 | 0.105 | 0.117 | 0.152 | 0.240 | 0.259 | 0.307 | 0.522 |
| 0.093 | 0.081 | 0.121 | 0.145 | 0.163 | 0.198 | 0.223 | 0.303 |
| 0.022 | 0.054 | 0.098 | 0.138 | 0.199 | 0.253 | 0.269 | 0.39 |

Table 21.9. Irish Sea plaice: Estimated landed numbers-at-age (thousands).

| year\age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 23 | 1720 | 4565 | 3179 | 829 | 247 | 108 | 250 | 10919 |
| 1982 | 23 | 871 | 3798 | 3316 | 1331 | 360 | 130 | 216 | 10044 |
| 1983 | 49 | 2676 | 2749 | 2939 | 1438 | 641 | 205 | 229 | 10927 |
| 1984 | 45 | 2709 | 4943 | 1436 | 1088 | 534 | 294 | 231 | 11280 |
| 1985 | 5 | 2459 | 5190 | 3082 | 613 | 508 | 272 | 310 | 12438 |
| 1986 | 35 | 1851 | 5328 | 3091 | 1438 | 298 | 290 | 348 | 12680 |
| 1987 | 59 | 3199 | 4886 | 3719 | 1475 | 784 | 160 | 406 | 14686 |
| 1988 | 56 | 2664 | 4984 | 2594 | 1385 | 611 | 347 | 316 | 12958 |
| 1989 | 33 | 1704 | 4896 | 2601 | 917 | 503 | 234 | 307 | 11194 |
| 1990 | 12 | 1156 | 2978 | 3142 | 1134 | 411 | 237 | 270 | 9340 |
| 1991 | 150 | 1604 | 1758 | 1912 | 1500 | 525 | 195 | 266 | 7911 |
| 1992 | 143 | 2072 | 3167 | 1196 | 1041 | 907 | 318 | 301 | 9145 |
| 1993 | 32 | 1239 | 2113 | 1721 | 402 | 376 | 411 | 285 | 6577 |
| 1994 | 96 | 1187 | 2926 | 1446 | 710 | 205 | 168 | 330 | 7068 |
| 1995 | 26 | 1005 | 1763 | 2087 | 619 | 289 | 95 | 225 | 6108 |
| 1996 | 24 | 982 | 1384 | 1222 | 999 | 304 | 125 | 165 | 5206 |
| 1997 | 25 | 664 | 1844 | 1305 | 722 | 566 | 199 | 193 | 5518 |
| 1998 | 6 | 716 | 1173 | 1097 | 664 | 382 | 308 | 250 | 4597 |
| 1999 | 58 | 754 | 1347 | 1226 | 624 | 317 | 186 | 255 | 4766 |
| 2000 | 376 | 401 | 1006 | 1563 | 780 | 292 | 175 | 227 | 4820 |
| 2001 | 14 | 462 | 888 | 1101 | 1100 | 369 | 145 | 222 | 4301 |
| 2002 | 1 | 228 | 1133 | 1028 | 834 | 636 | 195 | 210 | 4264 |
| 2003 | 208 | 280 | 1008 | 1246 | 645 | 428 | 286 | 225 | 4328 |
| 2004 | 19 | 259 | 795 | 926 | 747 | 251 | 160 | 236 | 3393 |
| 2005 | 9 | 223 | 733 | 1353 | 640 | 526 | 152 | 290 | 3926 |
| 2006 | 2 | 177 | 789 | 567 | 783 | 433 | 241 | 297 | 3290 |
| 2007 | 0 | 56 | 293 | 680 | 485 | 328 | 189 | 251 | 2283 |
| 2008 | 0 | 91 | 402 | 473 | 485 | 250 | 142 | 139 | 1983 |
| 2009 | 0 | 10 | 179 | 388 | 326 | 319 | 99 | 201 | 1522 |
| 2010 | 0 | 5 | 72 | 300 | 183 | 212 | 96 | 154 | 1022 |
| 2011 | 2 | 83 | 318 | 409 | 567 | 308 | 166 | 274 | 2127 |
| 2012 | 0 | 5 | 120 | 256 | 264 | 322 | 133 | 278 | 1378 |
| 2013 | 1 | 17 | 74 | 174 | 226 | 190 | 144 | 301 | 1128 |
| 2014 | 1 | 6 | 58 | 172 | 151 | 137 | 63 | 266 | 853 |
| 2015 | 0 | 1 | 39 | 97 | 165 | 118 | 114 | 382 | 915 |
| 2016 | 9 | 29 | 87 | 189 | 269 | 285 | 169 | 398 | 1436 |
| 2017 | 7 | 5 | 27 | 140 | 182 | 150 | 131 | 393 | 1036 |
| 2018 | 0 | 3 | 101 | 122 | 264 | 310 | 178 | 539 | 1518 |

Table 21.10. Irish Sea plaice: Estimated discarded numbers-at-age (thousands). All discards are included (dead and alive portions).

| year\age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 480 | 4531 | 5851 | 402 | 7 | 0 | 0 | 0 | 6762 |
| 1982 | 653 | 3129 | 3537 | 403 | 14 | 0 | 0 | 0 | 4642 |
| 1983 | 699 | 3451 | 1606 | 317 | 17 | 1 | 0 | 0 | 3654 |
| 1984 | 581 | 2759 | 1880 | 124 | 12 | 1 | 0 | 0 | 3214 |
| 1985 | 577 | 2684 | 1502 | 217 | 5 | 1 | 0 | 0 | 2991 |
| 1986 | 566 | 3034 | 1587 | 239 | 13 | 0 | 0 | 0 | 3264 |
| 1987 | 631 | 2549 | 1762 | 316 | 16 | 1 | 0 | 0 | 3165 |
| 1988 | 435 | 3557 | 2058 | 253 | 20 | 1 | 0 | 0 | 3796 |
| 1989 | 326 | 1951 | 2232 | 283 | 18 | 2 | 0 | 0 | 2887 |
| 1990 | 493 | 1450 | 1533 | 432 | 34 | 2 | 0 | 0 | 2366 |
| 1991 | 303 | 2687 | 1144 | 345 | 62 | 4 | 0 | 0 | 2728 |
| 1992 | 500 | 2248 | 2626 | 285 | 58 | 9 | 1 | 0 | 3435 |
| 1993 | 526 | 4339 | 2211 | 454 | 25 | 4 | 1 | 0 | 4536 |
| 1994 | 404 | 2950 | 3509 | 391 | 50 | 2 | 0 | 0 | 4384 |
| 1995 | 618 | 2618 | 2508 | 605 | 48 | 4 | 0 | 0 | 3841 |
| 1996 | 787 | 3541 | 2397 | 426 | 99 | 6 | 0 | 0 | 4353 |
| 1997 | 842 | 5925 | 3571 | 534 | 83 | 13 | 1 | 0 | 6581 |
| 1998 | 746 | 6879 | 6517 | 1461 | 234 | 25 | 4 | 0 | 9520 |
| 1999 | 495 | 4189 | 6669 | 1644 | 222 | 20 | 3 | 0 | 7945 |
| 2000 | 0 | 3350 | 4217 | 1897 | 220 | 14 | 2 | 0 | 5820 |
| 2001 | 517 | 3621 | 3182 | 1221 | 267 | 15 | 1 | 0 | 5295 |
| 2002 | 471 | 3763 | 3819 | 1081 | 186 | 24 | 1 | 0 | 5607 |
| 2003 | 0 | 3216 | 3605 | 1539 | 186 | 20 | 1 | 0 | 5140 |
| 2004 | 216 | 1805 | 2935 | 1408 | 344 | 19 | 1 | 0 | 4037 |
| 2005 | 283 | 2774 | 2892 | 1260 | 1001 | 89 | 10 | 35 | 5007 |
| 2006 | 464 | 2729 | 3772 | 2029 | 548 | 212 | 13 | 3 | 5862 |
| 2007 | 644 | 4134 | 4092 | 2821 | 907 | 312 | 131 | 194 | 7941 |
| 2008 | 419 | 4072 | 3323 | 1647 | 993 | 200 | 27 | 325 | 6604 |
| 2009 | 243 | 2208 | 3792 | 1369 | 628 | 294 | 60 | 101 | 5217 |
| 2010 | 506 | 3053 | 3956 | 3611 | 1708 | 701 | 502 | 479 | 8709 |
| 2011 | 387 | 2089 | 1813 | 1016 | 692 | 247 | 68 | 217 | 3917 |
| 2012 | 345 | 3135 | 2790 | 1540 | 701 | 498 | 201 | 337 | 5728 |
| 2013 | 239 | 1731 | 2619 | 1396 | 614 | 208 | 121 | 90 | 4211 |
| 2014 | 298 | 1900 | 2390 | 1925 | 1085 | 564 | 244 | 366 | 5263 |
| 2015 | 103 | 1247 | 1495 | 1153 | 800 | 558 | 130 | 102 | 3353 |
| 2016 | 66 | 647 | 1291 | 971 | 517 | 300 | 139 | 156 | 2452 |
| 2017 | 65 | 783 | 1210 | 1326 | 1061 | 694 | 332 | 547 | 3611 |
| 2018 | 108 | 707 | 994 | 909 | 689 | 410 | 187 | 216 | 2533 |

Table 21.11. Irish Sea plaice: Estimated population numbers-at-age (thousands).

| year\age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 16495 | 19801 | 17653 | 7241 | 2019 | 695 | 329 | 763 | 64996 |
| 1982 | 22526 | 12567 | 13234 | 7622 | 3261 | 1010 | 396 | 655 | 61272 |
| 1983 | 24055 | 20521 | 7903 | 6336 | 3292 | 1657 | 571 | 637 | 64972 |
| 1984 | 23097 | 21236 | 14271 | 3343 | 2713 | 1488 | 880 | 690 | 67718 |
| 1985 | 21331 | 20418 | 14644 | 7198 | 1534 | 1406 | 808 | 919 | 68259 |
| 1986 | 22025 | 17914 | 14778 | 7012 | 3450 | 778 | 816 | 976 | 67749 |
| 1987 | 21540 | 19783 | 12400 | 7519 | 3117 | 1791 | 396 | 1006 | 67552 |
| 1988 | 15728 | 20382 | 13224 | 5347 | 2930 | 1377 | 842 | 765 | 60596 |
| 1989 | 12442 | 13496 | 14656 | 5974 | 2155 | 1257 | 624 | 818 | 51420 |
| 1990 | 15931 | 9327 | 9220 | 7453 | 2730 | 1052 | 647 | 737 | 47098 |
| 1991 | 16174 | 14284 | 5694 | 4610 | 3634 | 1359 | 540 | 734 | 47029 |
| 1992 | 17719 | 12847 | 9593 | 2561 | 2148 | 1986 | 742 | 700 | 48297 |
| 1993 | 15629 | 16188 | 7682 | 4088 | 895 | 876 | 1024 | 708 | 47089 |
| 1994 | 14814 | 12325 | 11355 | 3529 | 1637 | 495 | 429 | 843 | 45427 |
| 1995 | 17498 | 10975 | 7643 | 5472 | 1556 | 773 | 271 | 638 | 44825 |
| 1996 | 22024 | 13608 | 6904 | 3542 | 2813 | 920 | 403 | 532 | 50744 |
| 1997 | 22871 | 18092 | 9637 | 3892 | 2071 | 1743 | 650 | 630 | 59586 |
| 1998 | 19684 | 20909 | 12279 | 4756 | 2187 | 1228 | 1041 | 837 | 62921 |
| 1999 | 18741 | 16989 | 15333 | 6370 | 2507 | 1276 | 811 | 1100 | 63127 |
| 2000 | 23790 | 14840 | 11792 | 9048 | 3556 | 1398 | 940 | 1212 | 66575 |
| 2001 | 24108 | 18898 | 10634 | 6885 | 5526 | 2003 | 919 | 1400 | 70372 |
| 2002 | 24726 | 20771 | 14985 | 7082 | 4649 | 3963 | 1473 | 1587 | 79235 |
| 2003 | 21854 | 22082 | 16466 | 10801 | 4378 | 3241 | 2762 | 2170 | 83754 |
| 2004 | 20535 | 17585 | 17288 | 11435 | 7160 | 2571 | 2169 | 3173 | 81915 |
| 2005 | 17721 | 18432 | 13042 | 10771 | 7357 | 4576 | 1783 | 3509 | 77190 |
| 2006 | 22571 | 15203 | 14689 | 8556 | 6307 | 4315 | 2867 | 3427 | 77935 |
| 2007 | 26478 | 18661 | 12021 | 10531 | 5435 | 3708 | 2970 | 4065 | 83869 |
| 2008 | 21204 | 23182 | 13210 | 8270 | 7259 | 3363 | 2228 | 4685 | 83400 |
| 2009 | 17106 | 16699 | 18260 | 9095 | 6138 | 5762 | 2422 | 4700 | 80183 |
| 2010 | 23108 | 15703 | 12816 | 13001 | 7263 | 4964 | 4814 | 5340 | 87010 |
| 2011 | 27130 | 17758 | 11415 | 8210 | 8910 | 5232 | 3655 | 7121 | 89432 |
| 2012 | 23650 | 24832 | 14382 | 9315 | 6010 | 6697 | 4215 | 7970 | 97071 |
| 2013 | 23617 | 20486 | 19635 | 11926 | 7804 | 4976 | 5285 | 8633 | 102362 |
| 2014 | 28211 | 21940 | 17647 | 15223 | 9915 | 6776 | 4561 | 10580 | 114853 |
| 2015 | 16576 | 24085 | 17998 | 13904 | 11485 | 8642 | 5409 | 12446 | 110546 |
| 2016 | 14079 | 15087 | 19168 | 15679 | 11699 | 9782 | 7636 | 14830 | 107959 |
| 2017 | 9902 | 13200 | 13306 | 15297 | 13460 | 9709 | 8111 | 17737 | 100721 |
| 2018 | 11443 | 10154 | 11389 | 10770 | 11650 | 10471 | 8031 | 18447 | 92355 |

Table 21.12. Irish Sea plaice: Estimated fishing mortality-at-age.

| year\age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | Fbar (3-6) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.020 | 0.272 | 0.660 | 0.689 | 0.571 | 0.469 | 0.421 | 0.421 | 0.597 |
| 1982 | 0.020 | 0.264 | 0.640 | 0.678 | 0.567 | 0.471 | 0.425 | 0.425 | 0.589 |
| 1983 | 0.021 | 0.282 | 0.685 | 0.736 | 0.623 | 0.524 | 0.473 | 0.473 | 0.642 |
| 1984 | 0.018 | 0.246 | 0.597 | 0.647 | 0.555 | 0.476 | 0.433 | 0.433 | 0.569 |
| 1985 | 0.018 | 0.238 | 0.579 | 0.637 | 0.550 | 0.480 | 0.437 | 0.437 | 0.562 |
| 1986 | 0.018 | 0.246 | 0.596 | 0.667 | 0.583 | 0.518 | 0.468 | 0.468 | 0.591 |
| 1987 | 0.022 | 0.293 | 0.705 | 0.793 | 0.696 | 0.619 | 0.550 | 0.550 | 0.703 |
| 1988 | 0.022 | 0.288 | 0.686 | 0.779 | 0.698 | 0.631 | 0.568 | 0.568 | 0.699 |
| 1989 | 0.020 | 0.257 | 0.597 | 0.672 | 0.605 | 0.549 | 0.501 | 0.501 | 0.606 |
| 1990 | 0.021 | 0.262 | 0.592 | 0.657 | 0.590 | 0.532 | 0.486 | 0.486 | 0.593 |
| 1991 | 0.022 | 0.273 | 0.604 | 0.664 | 0.591 | 0.526 | 0.479 | 0.479 | 0.596 |
| 1992 | 0.027 | 0.332 | 0.738 | 0.826 | 0.749 | 0.661 | 0.598 | 0.598 | 0.743 |
| 1993 | 0.024 | 0.290 | 0.641 | 0.722 | 0.675 | 0.605 | 0.549 | 0.549 | 0.660 |
| 1994 | 0.025 | 0.294 | 0.631 | 0.698 | 0.647 | 0.579 | 0.530 | 0.530 | 0.639 |
| 1995 | 0.025 | 0.287 | 0.601 | 0.639 | 0.578 | 0.507 | 0.461 | 0.461 | 0.581 |
| 1996 | 0.024 | 0.278 | 0.566 | 0.581 | 0.506 | 0.435 | 0.396 | 0.396 | 0.522 |
| 1997 | 0.025 | 0.285 | 0.575 | 0.581 | 0.500 | 0.426 | 0.389 | 0.389 | 0.521 |
| 1998 | 0.025 | 0.282 | 0.576 | 0.577 | 0.491 | 0.417 | 0.377 | 0.377 | 0.515 |
| 1999 | 0.021 | 0.229 | 0.461 | 0.458 | 0.384 | 0.317 | 0.279 | 0.279 | 0.405 |
| 2000 | 0.017 | 0.190 | 0.382 | 0.380 | 0.316 | 0.257 | 0.220 | 0.220 | 0.334 |
| 2001 | 0.015 | 0.161 | 0.327 | 0.331 | 0.276 | 0.222 | 0.183 | 0.183 | 0.289 |
| 2002 | 0.012 | 0.136 | 0.278 | 0.289 | 0.242 | 0.191 | 0.150 | 0.150 | 0.250 |
| 2003 | 0.010 | 0.113 | 0.229 | 0.239 | 0.202 | 0.155 | 0.116 | 0.116 | 0.206 |
| 2004 | 0.008 | 0.085 | 0.171 | 0.179 | 0.152 | 0.114 | 0.082 | 0.082 | 0.154 |
| 2005 | 0.011 | 0.115 | 0.224 | 0.233 | 0.197 | 0.144 | 0.098 | 0.098 | 0.199 |
| 2006 | 0.013 | 0.136 | 0.249 | 0.250 | 0.206 | 0.148 | 0.096 | 0.096 | 0.213 |
| 2007 | 0.016 | 0.156 | 0.278 | 0.273 | 0.224 | 0.159 | 0.100 | 0.100 | 0.233 |
| 2008 | 0.013 | 0.124 | 0.214 | 0.207 | 0.172 | 0.124 | 0.078 | 0.078 | 0.179 |
| 2009 | 0.009 | 0.089 | 0.154 | 0.152 | 0.129 | 0.095 | 0.061 | 0.061 | 0.133 |
| 2010 | 0.014 | 0.133 | 0.226 | 0.224 | 0.194 | 0.145 | 0.091 | 0.091 | 0.197 |
| 2011 | 0.009 | 0.083 | 0.140 | 0.141 | 0.124 | 0.097 | 0.062 | 0.062 | 0.126 |
| 2012 | 0.009 | 0.084 | 0.142 | 0.144 | 0.128 | 0.103 | 0.066 | 0.066 | 0.129 |
| 2013 | 0.007 | 0.056 | 0.093 | 0.094 | 0.084 | 0.069 | 0.044 | 0.044 | 0.085 |
| 2014 | 0.007 | 0.057 | 0.094 | 0.097 | 0.090 | 0.077 | 0.050 | 0.050 | 0.089 |
| 2015 | 0.004 | 0.034 | 0.057 | 0.062 | 0.061 | 0.057 | 0.038 | 0.038 | 0.059 |
| 2016 | 0.004 | 0.030 | 0.049 | 0.054 | 0.054 | 0.052 | 0.036 | 0.036 | 0.052 |
| 2017 | 0.005 | 0.039 | 0.062 | 0.067 | 0.067 | 0.064 | 0.044 | 0.044 | 0.065 |
| 2018 | 0.006 | 0.046 | 0.067 | 0.068 | 0.064 | 0.058 | 0.039 | 0.039 | 0.064 |

Table 21.13. Irish Sea plaice: SAM stock assessment summary ( $\pm 2$ standard deviation uncertainty). Recruitment ( 000 s ), spawning-stock biomass (SSB, tonnes), mean fishing mortality (Fbar) for ages 3-6, total spawning biomass (TSBS, tonnes), landings and discards tonnage.

| Year | Recruitment |  |  | SSB (t) |  |  | Fbar (3-6) |  |  | TSB (t) |  |  | Discards (t) |  |  | Landings (t) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | Mid | High | Low | Mid | High | Low | Mid | High | Low | Mid | High | Low | Mid | High | Low | Mid | High |
| 1981 | 11220 | 16495 | 24249 | 5751 | 7066 | 8681 | 0.462 | 0.597 | 0.773 | 10595 | 12799 | 15462 | 595 | 803 | 1084 | 2691 | 3632 | 4904 |
| 1982 | 16118 | 22526 | 31483 | 5572 | 6753 | 8184 | 0.465 | 0.589 | 0.747 | 10463 | 12436 | 14780 | 422 | 544 | 702 | 2721 | 3509 | 4525 |
| 1983 | 17331 | 24055 | 33387 | 5079 | 6066 | 7246 | 0.507 | 0.642 | 0.812 | 10888 | 12978 | 15469 | 324 | 409 | 517 | 2773 | 3506 | 4434 |
| 1984 | 16717 | 23097 | 31913 | 6406 | 7673 | 9191 | 0.449 | 0.569 | 0.72 | 13291 | 15897 | 19015 | 285 | 363 | 462 | 3298 | 4201 | 5352 |
| 1985 | 15469 | 21331 | 29415 | 7040 | 8443 | 10127 | 0.445 | 0.562 | 0.709 | 13906 | 16621 | 19865 | 263 | 336 | 430 | 3698 | 4727 | 6043 |
| 1986 | 15956 | 22025 | 30403 | 7533 | 9032 | 10829 | 0.47 | 0.591 | 0.743 | 14153 | 16788 | 19913 | 288 | 367 | 467 | 3959 | 5041 | 6420 |
| 1987 | 15499 | 21540 | 29936 | 7095 | 8450 | 10064 | 0.562 | 0.703 | 0.88 | 13997 | 16591 | 19667 | 285 | 361 | 457 | 4251 | 5383 | 6817 |
| 1988 | 11420 | 15728 | 21662 | 6614 | 7884 | 9398 | 0.557 | 0.699 | 0.876 | 12442 | 14715 | 17403 | 343 | 432 | 544 | 3781 | 4763 | 6000 |
| 1989 | 8815 | 12442 | 17561 | 5865 | 7033 | 8433 | 0.482 | 0.606 | 0.762 | 10793 | 12847 | 15292 | 267 | 341 | 436 | 3144 | 4020 | 5139 |
| 1990 | 11607 | 15931 | 21866 | 5327 | 6390 | 7666 | 0.472 | 0.593 | 0.745 | 9369 | 11069 | 13078 | 220 | 280 | 355 | 2726 | 3463 | 4401 |
| 1991 | 11864 | 16174 | 22048 | 4285 | 5103 | 6076 | 0.476 | 0.596 | 0.747 | 9124 | 10778 | 12730 | 249 | 312 | 391 | 2141 | 2681 | 3357 |
| 1992 | 13130 | 17718 | 23910 | 4285 | 5103 | 6077 | 0.599 | 0.743 | 0.923 | 8500 | 10012 | 11793 | 324 | 405 | 507 | 2422 | 3030 | 3791 |
| 1993 | 11913 | 15629 | 20505 | 3544 | 4233 | 5056 | 0.526 | 0.66 | 0.83 | 7530 | 8894 | 10505 | 413 | 517 | 646 | 1794 | 2244 | 2806 |
| 1994 | 11258 | 14814 | 19494 | 3653 | 4418 | 5343 | 0.51 | 0.639 | 0.8 | 7356 | 8726 | 10351 | 417 | 521 | 651 | 1842 | 2302 | 2877 |
| 1995 | 13336 | 17498 | 22958 | 3199 | 3884 | 4715 | 0.46 | 0.581 | 0.734 | 6673 | 7919 | 9399 | 362 | 453 | 566 | 1532 | 1915 | 2394 |
| 1996 | 16750 | 22024 | 28957 | 3414 | 4175 | 5104 | 0.411 | 0.522 | 0.663 | 7251 | 8637 | 10287 | 404 | 500 | 619 | 1425 | 1764 | 2182 |
| 1997 | 17425 | 22871 | 30018 | 3583 | 4377 | 5347 | 0.412 | 0.521 | 0.657 | 7783 | 9282 | 11071 | 608 | 753 | 934 | 1360 | 1686 | 2091 |
| 1998 | 15008 | 19684 | 25817 | 3864 | 4755 | 5851 | 0.402 | 0.515 | 0.66 | 8040 | 9629 | 11532 | 916 | 1140 | 1420 | 1206 | 1501 | 1869 |
| 1999 | 14189 | 18741 | 24754 | 4400 | 5484 | 6836 | 0.309 | 0.405 | 0.531 | 8880 | 10727 | 12957 | 791 | 985 | 1227 | 1220 | 1519 | 1892 |
| 2000 | 17720 | 23790 | 31940 | 4769 | 6015 | 7588 | 0.247 | 0.334 | 0.451 | 9074 | 11071 | 13507 | 587 | 737 | 926 | 1163 | 1461 | 1835 |


| Year | Recruitment (thousands) |  |  | SSB (t) |  |  | Fbar (3-6) |  |  | TSB (t) |  |  | Discards ( t ) |  |  | Landings (t) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | Mid | High | Low | Mid | High | Low | Mid | High | Low | Mid | High | Low | Mid | High | Low | Mid | High |
| 2001 | 18170 | 24108 | 31986 | 5712 | 7317 | 9373 | 0.212 | 0.289 | 0.395 | 10396 | 12809 | 15783 | 518 | 644 | 800 | 1262 | 1569 | 1951 |
| 2002 | 18514 | 24726 | 33023 | 6655 | 8646 | 11234 | 0.182 | 0.25 | 0.343 | 11528 | 14415 | 18023 | 549 | 680 | 843 | 1249 | 1548 | 1919 |
| 2003 | 16168 | 21854 | 29540 | 7827 | 10324 | 13617 | 0.146 | 0.206 | 0.291 | 12870 | 16374 | 20833 | 515 | 646 | 809 | 1224 | 1533 | 1920 |
| 2004 | 15321 | 20535 | 27523 | 7926 | 10492 | 13889 | 0.108 | 0.154 | 0.221 | 12595 | 16124 | 20642 | 452 | 573 | 725 | 894 | 1131 | 1432 |
| 2005 | 13279 | 17721 | 23650 | 7816 | 10306 | 13590 | 0.141 | 0.199 | 0.281 | 12502 | 15839 | 20065 | 568 | 713 | 893 | 1044 | 1309 | 1642 |
| 2006 | 16999 | 22571 | 29970 | 7035 | 9323 | 12354 | 0.153 | 0.213 | 0.297 | 12399 | 15617 | 19671 | 644 | 802 | 1000 | 884 | 1101 | 1372 |
| 2007 | 19740 | 26478 | 35514 | 5803 | 7704 | 10227 | 0.168 | 0.233 | 0.324 | 8897 | 11330 | 14429 | 719 | 898 | 1121 | 609 | 761 | 951 |
| 2008 | 15946 | 21204 | 28196 | 5826 | 7708 | 10196 | 0.13 | 0.179 | 0.248 | 10183 | 12837 | 16184 | 623 | 773 | 960 | 514 | 637 | 791 |
| 2009 | 12669 | 17106 | 23098 | 6553 | 8745 | 11670 | 0.095 | 0.133 | 0.186 | 9223 | 11890 | 15328 | 494 | 624 | 788 | 390 | 492 | 621 |
| 2010 | 17330 | 23108 | 30811 | 6622 | 8672 | 11355 | 0.141 | 0.197 | 0.277 | 10595 | 13351 | 16823 | 1038 | 1322 | 1684 | 274 | 349 | 444 |
| 2011 | 20218 | 27130 | 36404 | 7675 | 10320 | 13878 | 0.09 | 0.126 | 0.176 | 11813 | 15187 | 19523 | 379 | 472 | 589 | 565 | 704 | 878 |
| 2012 | 17705 | 23650 | 31591 | 6416 | 8672 | 11720 | 0.093 | 0.129 | 0.18 | 8442 | 11052 | 14469 | 407 | 506 | 630 | 362 | 450 | 560 |
| 2013 | 17673 | 23617 | 31561 | 7734 | 10471 | 14176 | 0.061 | 0.085 | 0.119 | 11193 | 14547 | 18906 | 338 | 421 | 525 | 305 | 380 | 473 |
| 2014 | 20429 | 28211 | 38959 | 7852 | 10576 | 14245 | 0.064 | 0.089 | 0.125 | 11314 | 14681 | 19051 | 430 | 535 | 666 | 265 | 331 | 412 |
| 2015 | 12140 | 16576 | 22635 | 10143 | 14118 | 19649 | 0.042 | 0.059 | 0.083 | 13365 | 17907 | 23993 | 237 | 298 | 375 | 325 | 409 | 514 |
| 2016 | 10452 | 14079 | 18964 | 14898 | 20433 | 28023 | 0.037 | 0.052 | 0.074 | 19112 | 25409 | 33781 | 242 | 303 | 379 | 509 | 638 | 799 |
| 2017 | 7099 | 9902 | 13812 | 11448 | 15622 | 21319 | 0.046 | 0.065 | 0.091 | 14294 | 18972 | 25181 | 359 | 449 | 562 | 354 | 443 | 554 |
| 2018 | 7696 | 11443 | 17014 | 12572 | 17522 | 24422 | 0.045 | 0.064 | 0.092 | 14128 | 19350 | 26502 | 206 | 265 | 340 | 477 | 613 | 789 |

Table 21.14 Short term forecast. Annual catch options. All weights are in tonnes.

|  |  |  | Total unwanted |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total catch (2020) | Wanted catch (2020) | $\begin{aligned} & \text { catch }^{*} \\ & (2020) \end{aligned}$ | F total (2020) | $\begin{gathered} \text { F wanted } \\ (2020) \end{gathered}$ | $\begin{gathered} \text { F unwanted** } \\ (2020) \end{gathered}$ | $\begin{gathered} \text { SSB } \\ (2021) \end{gathered}$ | $\begin{gathered} \text { dSSB } \\ * * * \end{gathered}$ | dTAC^ | dAdvice ${ }^{\wedge \wedge}$ |
| Fmsy | 3299 | 1931 | 1368 | 0.196 | 0.054 | 0.142 | 18354 | -4 | 7 | -6 |
| $\mathrm{F}=\mathrm{FMSY}$ lower | 2288 | 1341 | 947 | 0.133 | 0.036 | 0.097 | 16576 | -13 | -26 | -35 |
| $\mathrm{F}=\mathrm{FMSY}$ upper | 4772 | 2779 | 1993 | 0.293 | 0.08 | 0.213 | 17057 | -11 | 55 | 36 |
| $\mathrm{F}=0$ | 0 | 0 | 0 | 0 | 0 | 0 | 21332 | 11 | -100 | -100 |
| $\mathrm{F}=\mathrm{Fpa}$ | 5640 | 3300 | 2340 | 0.355 | 0.097 | 0.258 | 16371 | -15 | 83 | 61 |
| $\mathrm{F}=\mathrm{Flim}$ | 7467 | 4375 | 3092 | 0.495 | 0.136 | 0.359 | 14812 | -23 | 143 | 113 |
| SSB (2021) = Blim | 20103 | 12253 | 7850 | 2.515 | 0.689 | 1.826 | 3958 | -79 | 554 | 474 |
| SSB (2021)=B pa | 18728 | 11285 | 7443 | 2.063 | 0.565 | 1.498 | 5294 | -72 | 509 | 435 |
| SSB (2021)= MSY B | 14839 | 8839 | 6000 | 1.287 | 0.353 | 0.934 | 8757 | -54 | 383 | 324 |
| $\mathrm{F}=\mathrm{F} 2019$ | 1070 | 625 | 445 | 0.06 | 0.017 | 0.043 | 20323 | 6 | -65 | -69 |
| * Dead + surviving unwanted catch. |  |  |  |  |  |  |  |  |  |  |
| **F unwanted concerns dead unwanted catch only. |  |  |  |  |  |  |  |  |  |  |
| *** SSB 2021 relative to SSB 2020. |  |  |  |  |  |  |  |  |  |  |
| ${ }^{\wedge}$ Total catch in 2020 relative to TAC in 2019 (3075 t). |  |  |  |  |  |  |  |  |  |  |
| $\wedge \wedge$ Catch advice value 2020 relative to catch advice value 2019 ( $\mathbf{3 5 0 3} \mathbf{t}$ ). |  |  |  |  |  |  |  |  |  |  |



Figure 21.1. Irish Sea plaice: Effort and lpue for commercial fleets from UK (E\&W), Ireland and Belgium.


Figure 21.2a. Landings-at-age data (left) and mean standardised proportion-at-age (right, black bubbles are positive values and orange bubbles are negative). Mean standardised proportion-at-age $=$ [ (proportion-at-age in year) - mean (proportion-at-age over all years) ] / STDEV(proportion-at-age over all years).


Figure 21.2b. Discards-at-age data (left) and mean standardised proportion-at-age (right, black bubbles are positive values and orange bubbles are negative). Mean standardised proportion-at-age $=[$ (proportion-at-age in year) - mean (proportion-at-age over all years) ] / STDEV(proportion-at-age over all years).



Figure 21.4. Length distributions of discarded and retained catches of year 2018 from UK(E\&W) as available at InterCatch.



Figure 21.5. Length distributions of discarded and retained catches year 2018 from Ireland as available at InterCatch.


Figure 21.6. Northern Irish Groundfish Survey SSB indices split into spring (left hand panels) and autumn (right hand panels) sampling by western strata (1-3), eastern strata (4-7) and total survey area (strata 1-7) with confidence intervals ( $\pm 1$ standard error, vertical lines).


Figure 21.7. Trends in biomass indices ( kg per km towed) the NIGFS-WIBTS-Q1 and -Q4 (blue and red lines respectively) in the eastern Irish Sea (top) and the western and southern Irish Sea (bottom). Also shown (green dots, right axis) are the estimates of SSB from the Annual Egg Production Method (AEPM) from Armstrong et al. (2011).


Figure 21.8. Selectivity of the fishery split into the landed (solid) and discarded (dashed) components as estimated by the SAM model, where the $x$-axis shows age and the $y$-axis gives the fishing mortality-at-age scaled so that the maximum value is 1 and split by the proportion of fish (by number) discarded and landed at-age.


Figure 21.9. Catchability for the UK (E\&W)-BTS-Q3 extended index by age, NIGFS-WIBTS-Q1 and NIGFS-WIBTS-Q4 as estimated by the SAM model.


Figure 21.10. Residuals in fits to catch and survey data from the baseline model. Expected values were estimated by the SAM model.


Figure 21.11. Comparison of the standardised age 1 index from the UK (E\&W)-BTS-Q3 extended area (red) and the standardised recruitment (blue dashed line) estimated by the SAM model.


Figure 21.12. SAM model estimates of mean standardised SSB (orange line) overlain with standardised NIGFS in spring (red) and autumn (blue dashed) relative SSB indices, standardised biomass (ages 1-4) from the UK(E\&W)-BTS (black solid line) and AEPM SSB index (circles, right axis). Standardized: minus mean and divided by standard deviation.


Figure 21.13. Modelled SSB (tonnes, top left), recruitment (thousands, bottom left), F bar $^{\text {(ages 3-6, bottom right) }}$ catch tonnage (bottom right) using the SAM model. Error dashed lines indicate $2 \times$ standard deviation.


Figure 21.14. Retrospective assessments for years 2008-2018 from the baseline model. SSB (tonnes, top left), recruitment (thousands, bottom left), $\mathrm{F}_{\text {bar }}$ (ages 3-6, bottom right) catch tonnage (bottom right). Error dashed lines indicate $2 \times$ standard deviation.

# 23 Plaice (Pleuronectes platessa) in divisions 7.b-c (West of Ireland) 

## Type of assessment in 2019

No assessment was performed.

### 23.1 General

## Stock Identity

Plaice in 7.b are mainly caught by Irish vessels on sandy grounds in coastal areas. Plaice catches in 7.c are negligible. There are two distinct areas in which plaice are caught by Irish vessels in 7.b: an area around Galway Bay and an area in the north of 7.b, which extends into 6.a (the Stags and Broadhaven Ground). During 1995-2000 a large proportion of the 7.bc plaice landings were taken from the Stags Grounds (Rectangles 37D8, 37D9, 37E0 and 37E1). The landings and LPUE in this area have dropped sharply since 2000, in line with a general decrease of LPUE in Division 6.a. Plaice in this area appear to be more linked with $6 . a$ than populations further south. The landings and LPUE on the Aran grounds appear to have been more or less stable since the start of the logbooks' time-series in 1995 (WD 1,WGCSE 2009). It is not known, how much exchange there is between plaice on the Aran grounds and those on the Stags ground. The commercial LPUE time-series may not be reflective of overall stock abundance due to changing fishing practices.

## Data

The time-series of official landings is presented in Table 22.1 and Figure 22.1.
Sampling is carried out in Ireland but numbers of samples varies over time due to the low landings levels and varying encounter probability and is not sufficient to generate a time-series of annual length or age distributions. Sampling in 2018 was relatively good with 13 length sample units (1289 fish measured) and seven discard trips. Figure 22.2 describes the length-frequency distribution of the discard trips, and the contribution of these length classes to hauls and trips, no landings were recorded on these trips in 2018.

Table 22.1. Landings of plaice in 7.bc as officially reported to ICES.




Figure 22.1. Landings of plaice in 7.bc as officially reported to ICES (1908-2018).


Figure 22.2. Estimated age distribution of plaice 7.bc in 2018 based on Irish sampling (landings in blue, discards in black).

## 24 Plaice (Pleuronectes platessa) in Division 7.e (western English Channel

## Type of assessment in 2018

Last year's assessment report is available at: http://www.ices.dk/sites/pub/Publication\ Reports/Expert\ Group\ Report/acom/2018/WGCSE/27_Plaice_7e_2018.pdf

## ICES advice applicable to 2019

Last year's advice is available at: http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2018/2018/ple.27.7e.pdf

### 24.1 General

## Stock description and management units

The management area for this stock is strictly that for ICES Division 7.e, called the Western English Channel. The TAC area does not correspond to the stock area given that it includes the larger component of 7.d (Eastern English Channel). However, WKFLAT 2010 found that a significant proportion of the catches of the 7.e stock are taken in the adjacent division during the spawning period. Plaice is not the main target species in $7 . e$, and it is generally taken as bycatch in fisheries targeting sole.


TAC area $=7 . \mathrm{d}-\mathrm{e}$; Assessment area $=$ 7.e.

Management applicable to 2017 and 2018
There are technical measures in operation including a minimum 80 mm mesh size and a minimum landings size $(27 \mathrm{~cm})$ for this species.

The TAC and the national quotas by country for 2018

| Species: | Plaice <br> Pleuronectes platessa | Zone: | 7d and 7e <br> (PLE/7DE.) |
| :--- | :--- | :--- | :--- |
| Belgium | 1695 |  |  |
| France | 5651 |  |  |
| United Kingdom | 3014 | Analytical TAC |  |
| Union | 10360 |  |  |
| TAC | 10360 |  |  |

(Source: Council Regulation (EU) 2018/120, EU, 2018b).

The TAC and the national quotas by country for 2019

| Species: | Plaice <br> Pleuronectes platessa | Zone: <br> 7d and 7e <br> (PLE/7DE.) |
| :--- | :--- | :--- |
| Belgium | 1694 |  |
| France | 5648 |  |
| United Kingdom | 3012 | Analytical TAC |
| Union | 10354 | Article 7(2) of this Regulation applies |
| TAC | 10354 |  |

(Source: Council Regulation (EU) 2019/124, EU, 2019).

## Landing obligation

The landing obligation is being phased in between 2019 and 2021 for plaice in $7 . e$ with a discard plan defined in the Commission Delegated Regulation (EU) 2018/2034 (EU, 2018a) and referring to Regulation (EU) No 1380/2013 (EU, 2013). According to this discard plan, the landing obligation applies to plaice in 7.e. since 1 January 2019. There are, however, survivability exemptions for plaice when caught with specific gears. This includes all (a) trammelnets and (b) otter trawls. Furthermore, there is a provisional exemption for only 2019 including BT2 beam trawls (i.e. 80 mm to 120 mm mesh size) for (c) vessels with a maximum engine power greater than 221 kW and fitted with a flip-up rope or benthic release panel, and (d) for vessels with a maximum engine power of 221 kW or a maximum length of 24 m , when fishing within 12 nautical miles of the coast and with average tow durations of no more than 1:30 hours. (Commission Delegated Regulation (EU) 2018/2034, Article 6, EU, 2018a).

Prior to the introduction of the landing obligation, a substantial part of the plaice 7.e catches has been discarded and not accounted for in the stock assessment. In the first year of the phasing in of the landing obligation, the exemptions are likely to cover the majority of plaice catches and the impact on fishing or stock assessment is likely to be negligible. In the following two years of the discard plan, the situation should be closely monitored because of potential changes in the landings data and composition which might affect the stock assessment.

### 24.2 Data

## The fishery in 2018

International catch data are collated on the ICES InterCatch platform. In the Western English Channel, plaice are taken mainly as bycatch in bottom trawls targeting sole and anglerfish. In $2018,70 \%$ of the landings were taken by beam trawls, $23 \%$ by otter trawls, $4.9 \%$ by gillnets and $1.62 \%$ by other gears. Of the total international landings $84 \%$ were taken by the UK, $7.1 \%$ by France, $8.7 \%$ by Belgium, $0.198 \%$ by Ireland and $0.20 \%$ by the Netherlands (Table 23.1, Figures 23.1 and 23.2).

This stock is the smaller of the two plaice stocks that make up the larger TAC Area 7d-e. The official landings from this stock amounted to $19 \%$ of the TAC in 2017 and $16 \%$ of the TAC in 2018.

## Landings

National landings data reported to ICES and estimates of total landings used by the Working Group are given in Table 23.1. Total international landings in for 7.e were 1915 t in 2017 and 1644 t in 2018.

In addition to the estimated 2018 landings for 7.e, an extra of 236 tonnes was added from the $7 . \mathrm{d}$ plaice stock representing an adjustment for migration of $15 \%$ of the mature component of quarter 1 landings between the two divisions. This process was agreed at WKFLAT 2010, and the migration correction was revised at WKPLE 2015. The process has been described in the Stock Annex. A reciprocal correction was made to the 7.d plaice stock at WGNSSK 2019. Figure 23.19 shows the total annual landings split by divisions 7.e and 7.d.

## Discards

Although discards have not been used in the assessment of 7.e plaice in the past, some discard data are available. Discard tonnages are available within InterCatch and were provided by the UK(E\&W) for the years 2012-2018, by France for 2014-2018, by Belgium for 2012-2013 and 20152018 and by Ireland for 2017-2018 (zero discards reported). Discard coverage and sampling is at a high level. In 2018, 469 t of discards were reported in InterCatch for 7 .e and these discard estimates accounted for $92 \%$ of reported landings. For the remaining $7.7 \%$ of landings for which no discard estimates were provided, an additional discard estimate of 21 t was raised. Age samples for discards have only been provided by the UK(E\&W) but cover the years 2012-2018. In 2018, the sampled discards covered $94 \%$ of the submitted discards.

In analogy to the landings, the discards are also uplifted by a migration correction from 7.d. For 2018, 236 t ( $15 \%$ of the mature Q1 plaice discards in 7.d) were added resulting in total discards of 624 t for the 7 .e plaice stock.

For historical reasons, Figure 23.3 shows various discard rates for plaice in 7.e. Since WGCSE 2017, the discard rate is calculated as the contribution of total discards (raised, including migration correction) to the total plaice landings (including migration component) and this discard rate was $25 \%$ in 2018 (Figure 23.4).

## Sampling

This year, all nations (apart from Scotland and the Belgian beam trawl fleet) provided data disaggregated by fleet and by quarter and these were all uploaded into the ICES InterCatch database. Quarterly age compositions for landings in 2018 were available only from the UK (England) only and were provided for six fleets (GNS_DEF_all_0_0_all Q1, Q2, Q3, Q4,

MIS_MIS_0_0_0_HC Q1,OTB_DEF_>=120_0_0_all Q2, Q3, OTB_DEF_70-99_0_0_all Q1, Q2, Q3, Q4, TBB_DEF_>=120_0_0_all Q4, TBB_DEF_70-99_0_0_all Q1, Q2, Q3, Q4, Figure 23.9). These data accounted for $80 \%$ of the total reported international landings. Additional landings data were available by quarter/fleet from Belgium, France, Ireland (Q3, Q4), the Netherlands, UK (E+W, Guernsey, Jersey) and UK Scotland (annual). These datasets were aggregated to an international age-structured catch using the ICES InterCatch platform.
Length compositions were provided by the UK(E\&W) and France (Figure 23.6).
An additional age composition representing the migration adjustment ( $15 \%$ of the mature component of quarter 1 landings for 7.d) was supplied on request by the WGNSSK stock coordinator for the 7.d plaice stock.

The method for the derivation of the international catch numbers and the calculation of the catch and stock weights-at-age has been fully described in the Stock Annex, Section B1 (Figure 23.17). Landings numbers-at-age (including the migration element) are given in Table 23.2 and in Figure 23.10. Landings and stock weights-at-age are given in Table 23.3 and Table 23.4 and plotted in Figure 23.18.
Catch weights are assumed to be mid-year values and stock weights are interpolated back (in year) to January 1st, as standard for this stock.

## Revisions

No revisions to data for previous years were submitted.

## Biological

The natural mortality and the maturity ogives used were identical to previous assessments and as described in the stock annex.

## Surveys

Two surveys currently provide abundanceestimates to the Working Group (Figure 23.14, Figure 23.15, Table 23.5 and Figure 23.16 for internal consistency).

In the 2018 FSP survey, landings numbers were dominated by plaice (34\%) and sole (27\%). In terms of total catch numbers (including discards), catches were dominated by bycatch species red gurnard and common dragonet (both $11 \%$ ), lesser spotted dogfish ( $9 \%$ ) and anglerfish ( $8 \%$ ). The actual targeted species sole and plaice accounted for $5 \%$ and $9 \%$ respectively of total catch numbers. Plaice continues to be widespread and were encountered at 82 of 89 stations with greater numbers caught in Bigbury Bay and Great West Bay and fewer in the western part of the survey area. The catch of commercially important species (sole, plaice, megrim, lemon sole, anglerfish, monkfish) has been dominated by plaice in recent years but has dropped in 2018 due to high catches of monkfish. Aggregated plaice cpue estimates for the UK FSP-7e survey increased continuously since 2009 and reached a time-series maximum in 2014 and have been declining again since then. Currently, ages 3, 4 and 5 are most abundant in the index. There are some year effects visible (high catches in 2014) but cohort effects are not evident from looking at the catch curves. The internal consistency is good for ages larger than three.

The Q1SWBeam survey is based on a stratified random survey approach that covers the entire region of the management area and some adjacent waters. Cpue estimates for the Q1SWBeam survey gradually increased from 2006 to 2012 and increased rapidly thereafter to reach the highest levels on record in 2014. Since then, the index is very variable, mainly caused by changes in the abundance of plaice aged three and four. The internal consistency is reasonable for ages three to six. There is no correlation between the catches-at-age two and three and between seven and eight.

## Commercial fleet effort and Ipue

UK (E\&W) beam trawl and otter trawl time-series are shown in Figure 23.5.
UK(E\&W) beam trawl effort is relatively stable at high levels since the early 2000s but the landings increased substantially in after 2015.

UK(E\&W) otter trawl effort (days fished-GRT corrected) has declined since 1989 to very low levels in recent years. In 2016, this fleet reported 0 effort and no landings, i.e. there is no lpue value for 2016. The reason for is that the lpue otter trawl index is calculated only with vessels of at least 12 m length and in 2016 only smaller vessels deploying otter trawls reported any activity. Due to a change in the database system, there are no consistent values available for the otter trawl fleet after 2016.

### 24.3 Data-limited methods

In 2017, ICES requested to trial data-limited methods for category 3 stocks in order to try to estimate the stock status relative to (proxy) MSY reference points and this is being provided since then.

During WGCSE 2019, the length-based indicator (LBI) method as developed during WKLIFE workshops (ICES, 2014;2015e) and theSurplusProduction in Continuous Time (SPiCT, Pedersen and Berg, 2017) were again applied.

## Length-based indicators

Internationally raised length frequencies from InterCatch were only available for 2014-2018. Figure 23.7 shows these length distributions (landings and discards), including the length of first capture $\left(L_{c}\right.$, calculated as first length class where the abundance is equal or larger to half of the maximum observed abundance in any length class) and the mean length in the catch above $L_{c}$.
As no country submitted life-history parameters for plaice, these had to be estimated with the available data. The data used were the age-length key from commercial UK data. For 2017, data were available for discards and landings, for 2018 only for landings. Von Bertalanffy growth functions were fitted to this dataset (Figure 23.8). This led to estimates of:

| data | $\boldsymbol{L}_{\text {inf }}[\mathrm{cm}]$ | $k\left[\right.$ year $\left.^{-1}\right]$ | $\boldsymbol{t}_{0}$ [years] |
| :--- | :--- | :---: | :---: |
| 2017 (landings and discards) | 67.570754 | 0.08029 | -3.86813 |
| 2018 | 81.31199 | 0.05747263 | -4.715559 |
| all | 69.05451 | 0.07841 | -3.86361 |

The estimates of $L_{\text {inf }}$ appear reasonable but the estimated growth rates $(k)$ seem too low for plaice. These estimates should be treated very carefully as they were based on commercial data and were from one country only. Furthermore, as the data are commercial, the data are most likely biased as fishery selectivity is not considered. The fish in the age-length key are mostly within the age range 4-7 and data for very young or older fish are scarce, impeding a robust model fit.

Several possible length-based indicators are suggested by ICES, but for plaice only the MSY proxy indicator $L_{\text {mean }} / L_{F=M}$ was calculated, as this is the only one that would be used to inform on the stock status regarding MSY proxies.

The reference length is defined as $L_{F=M}=\frac{L_{\text {inf }}+\frac{2 M}{l} L_{c}}{1+\frac{2 M}{k}}$.
Due to the uncertainty in the life-history parameters, the approximation of $M / k=1.5$ was adopted and $L_{i n f}$ used derived from all data combined. For the calculation of the proxy reference length, the life-history parameters or proxies were set constant for all years, whereas $L_{c}$ was calculated for every year. The following table shows the used parameters and the ratio $L_{\text {mean }} / L_{F=M}$ :

| year | 2014 | 2015 | 2016 | 2017 | 2018 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| $L_{C}[\mathrm{~cm}]$ | 24 | 26 | 26 | 27 | 26 |
| $L_{\text {mean }}[\mathrm{cm}]$ | 30.2 | 32.3 | 33.3 | 33.0 | 32.7 |
| $L_{F=M}[\mathrm{~cm}]$ | 35.3 | 36.8 | 36.8 | 37.5 | 36.8 |
| $L_{\text {mean }} / L_{F=M}$ | 0.857 | 0.878 | 0.906 | 0.879 | 0.889 |

According to this table, the stock would be slightly below the reference points for the last five years. At WGCSE 2019, these values were rejected (as in the previous years) due to uncertainty in in the life-history parameters, uncertainty about the assumption that $F=M$ is proxy for MSY and because a variety of data for this stock are available and falling back onto length data when a longer age-disaggregated catch history and survey data exist, is implausible.

## SPiCT

During WKProxy (ICES, 2016b) a SPiCT assessment with a set of predefined parameter settings was used to assess plaice in 7.e. The workshop came up with reference values for biomass (exploitable biomass) and fishing mortality and found that the stock is in a desirable state, both in term of biomass and fishing mortality. The results from this assessment are also available on www.stockassessment.org. This assessment was reviewed at WGCSE 2017 and rejected, as it could not be reproduced.

As in the last two years, a trial assessment with SPiCT was conducted for plaice in 7.e. The input data comprised the same two tuning surveys as used in the traditional XSA assessment (FSP-UK and Q1SWBeam). For both surveys, an annual biomass value was created by summing up the biomass catch-at-age for the same ages as used in XSA. The catch input for SPiCT consisted of the total landings time-series. Several attempts were made to fit SPiCT. The result from the baseline run are shown in Figure 23.11 and Table 23.11, the diagnostic plots in Figure 23.12.

Even though the model converged, the results are not appropriate. The uncertainty in the estimates were very high. This was true for the absolute and relative estimates as well as the estimated reference points. Furthermore, the datapoints on the production curve are all on the left, indicating a lack of contrast required for good model fit. According to the assessment estimates the stock is currently below $50 \%$ of $B_{\text {MSY }}$ and the fishing mortality well above FMSY, indicating an undesirable stock status.

The model was highly sensitive to the input data range. Figure 23.13 shows results of a retrospective and "inverted retrospective" (removing years from the beginning of the time-series) analysis. Removing years from the end or the beginning of the time-series led to a massive rescaling. This indicates a lack of consistency and the results of the model should be treated very cautiously.

Consequently, due to the high uncertainty, missing consistency and the inability of tracking stock dynamic properly the SPiCT model fit was rejected and should not be used to inform on the stock status.

As last year, instead of using proxy reference points, the stock status evaluation for plaice in 7.e is based on analytical reference points from an XSA assessment, as described later in this report.

### 24.4 Stock assessment

## Catch-at-age analysis

During this year's WGCSE an XSA assessment was performed with the settings defined in the stock annex.

## Data compilation and screening

The age range for the analysis was $2-10+$ in accordance with the updated procedures outlined at IBPWCFlat2 2015 and detailed in the Stock Annex. The landings data were processed according to the stock annex, and formed the reference dataset for this year's assessment.

As this was an update assessment, full data screening, tuning data and extensive exploratory XSA trials were not carried out.

Available tuning information consisted of five fleets: three UK commercial series, UK otter historic, UK otter trawl, UK beam trawl; and two UK survey series: FSP-7e (UK(E\&W)) and Q1SWBeam but in accordance with the decision of WGCSE in 2015, only the UK surveys were analysed and used in the assessment.

## Update assessment

The settings used for the final run are shown in the table below. The full assessment history is given in the stock annex.

|  |  | 2017 XSA | 2018 XSA | 2018 XSA |
| :---: | :---: | :---: | :---: | :---: |
| Catch-at-age data | Landings | 1980-2016, 2-10+, <br> $15 \%$ mature Q1 catch from 7.d added | 1980-2017, 2-10+, <br> 15\% mature Q1 catch from 7.d added | 1980-2018, 2-10+, <br> 15\% mature Q1 catch from 7.d added |
|  | Discards | - | - | - |
| Fleets | UK-WEC-BTS - Survey | - | - | - |
|  | UK WECOT - Commercial | - | - | - |
|  | UK WECOT-Commercial historic | - | - | - |
|  | UK WECBT - Commercial | - | - | - |
|  | FSP-7e - Survey | $\begin{gathered} \text { 2003-2016, 2-8 } \\ \text { (exc. 2008) } \end{gathered}$ | $\begin{gathered} \text { 2003-2017, 2-8 } \\ \text { (exc. 2008) } \end{gathered}$ | $\begin{gathered} \text { 2003-2018, 2-8 } \\ \text { (exc. 2008) } \end{gathered}$ |
|  | $\begin{aligned} & \text { Q1SWBeam - Sur- } \\ & \text { vey } \end{aligned}$ | 2006-2016, 2-9 | 2006-2017, 2-9 | 2006-2018, 2-9 |
| Taper |  | No | No | No |
| Taper range |  | - | - | - |
| Ages catch dep. Stock size |  | None | None | None |
| q plateau |  | 6 | 6 | 6 |
| F shrinkage se |  | 1.0 | 1.0 | 1.0 |
| Year range |  | 3 | 3 | 3 |
| Age range |  | 3 | 3 | 3 |
| Fleet SE threshold |  | 0.3 | 0.3 | 0.3 |
| Prior weighting |  | - | - | - |
| Plus group |  | 10 | 10 | 10 |
| F Bar Range |  | F(3-6) | F(3-6) | F(3-6) |

The log-catchability residuals for the XSA run (landings only) are shown in Figure 23.20. For 2016, most residuals for the UK-FSP survey are negative, whereas they are positive for the Q1SWBeam survey. This is because of contradictory signals from the two surveys. This behaviour did not repeat afterwards.

Fishing mortalities and stock numbers estimated from the final run are given in Table 23.6 and Table 23.7, and the assessment summary is shown in Table 23.8.

Retrospective patterns in stock status and fishing mortality estimates exhibited an unacceptably high degree of temporal variability since the late 1990s, thereby indicating an excessive level of uncertainty and a lack of robustness in the assessment outputs. Consequently, since 2015 the Working Group assessed the status of the plaice 7. estockusing a qualitative evaluation of survey trends only in accordance with the ICES Data-Limited Stock (DLS) category 3 approach.
A five-year retrospective analysis (Figure 23.21) was conducted. Compared to previous years, the retrospective has almost disappeared entirely and does not seem to be a problem anymore. The stock is still treated as a category 3 stock, mainly because of a missing discard time-series.

A Mohn's rho analysis was conducted based on the XSA stock assessment results, i.e. the last data year (2018) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (ages 3-6) | recruitment |
| :--- | :--- | :--- | :--- |
| Mohn's rho value | 0.00417205310419406 | -0.0149021219845442 | -0.127110812447642 |

The Mohn's rho values for this assessment are very low and well below the threshold of $20 \%$ imposed by ICES for 2019 assessments, i.e. the current assessment indicates a very high consistency.

## Comparison with previous assessments

The assessment shows a high consistency compared to last year's assessment (Figure 23.23).

## Intermediate year

As recommended by the advice drafting group (ADG) in 2017, the SSB is now presented until the intermediate year. All age groups used in the assessment contribute to the SSB (plaice at-age 2 , the recruitment age, are thought to be $26 \%$ mature). Therefore, to obtain an SSB estimate for 2018, an assumption about the recruitment in 2019 is required. Due to a lack of data to predict recruitment in 2019, the long-term (entire time-series) geometric mean was used.

## State of the stock

As in the last years, the XSA assessment based on landings data only was used as final assessment run. A summary of this assessment is given in Table 23.8 and Figure 23.22.

As this is a category 3 stock, the results of the stock assessment are indicative of trends only for giving advice. In Figure 23.22, absolute and relative (to mean of time-series) results of the stock assessment are presented. Table 23.8 shows the absolute assessment results and Table 23.9 relative assessment summary results.

Spawning-stock biomass was relatively stable from 1982 to 1985 and then increased until 1989 above the long-term average following strong recruitment events during the mid-1980s. Subsequently, spawning-stock biomass decreased until 1996. A strong year class in 1996 generated an
increase in spawning-stock biomassbetween 1996 and 2000 . However, successive poor year classes resulted in spawning-stock biomass declining to the lowest levels in 2007. A combination of above average recruitment and a reduction in fishing mortality has increased spawning-stock biomass since 2008 to reach the highest level on record in 2016. However, since then, the SSB has decreased but is still a high level.

Fishing mortality gradually increased from the 1980s up until the 2000s, peaking briefly in 2007. Following a large reduction in fishing mortality in 2009, this assessment shows a general decline that has reached the lowest levels on record in 2015. Since then, $F$ has increased again.

This assessment estimates that recruitment has been above the long-term geometric mean (19802018) between 2010 and 2015 and below afterwards.

However, the optimistic stock development in recent years is uncertain due to assessment uncertainty and omitting discard information. The decision to omit discard data is mainly due to uncertainty in the actual discard rate and unknown proportion of surviving plaice in the discards.

## State of the stock in comparison to analytical reference points

Analytical reference points for the landings only XSA assessment were estimated during WKMSYREF4 (ICES, 2016a) but never used due to the downgrading of the stock to category 3 . The main reason for this was an unacceptable retrospective pattern. This problem has now disappeared and consequently the analytical reference points have been restored in 2017.

In comparison with reference points, F surpassed $\mathrm{F}_{\text {MSY }}$ in 2016 and is above since then. The SSB is well above MSYB trigger $\mathrm{B}_{\mathrm{pa}}$ and $\mathrm{B}_{\text {lim }}$.

### 24.5 Short-term projections

As in the last years, plaice in 7.e continues to be treated as a category 3.2.0 stock and the assessment is indicative of trends only. Therefore, catch advice was provided by applying the ICES DLS framework for category 3 stocks where temporal trends in spawning-stock biomass are used as an index of stock development. The advice is based on a comparison of the two latest index values (index A) with the three preceding values (index B), multiplied by the recent advised catch. The SSB estimates from the landings only assessment are used as index values for this stock.

As basis for calculating the landings corresponding to the catch advice, the total catches as raised in InterCatch, including the migration correction from 7.d is used.

The basis for the catch options is presented in Table 23.10. For stocks in ICES data categories 36 , one catch option is provided.

The index ratio (2019-2018 / 2017-2015) is 0.93 and therefore the uncertainty cap does not need to be applied.

The fishing mortality derived from XSA is above FmSY $_{\text {M }}$ and the precautionary buffer has never been applied since this stock is treated as category 3, therefore the precautionary buffer should be applied and reduce the advised catch by $20 \%$.

The advised catch in 2018 for 2019 was 3648 tonnes and is used as the basis for calculating the new advice.

This leads to the following advice for the 7.e plaice stock:

ICES advises that when the precautionary approach is applied, catches of the Division 7.e plaice stock should be no more than 2721 tonnes in 2019. If discard rates do not change from the average of the last six years (2012-2018), this implies landings of no more than 1909 tonnes.

The average proportion of the landings of the 7.e plaice stock taken in Division 7.e between 20032018 is $10 \%$. The year range (2003 until most recent year) was agreed between the two stocks (7.e and 7.d) and is also used in the advice for the 7.d stock. The calculation of this proportion only includes landings and disregards discards, as discard estimates for the plaice 7.e stock only exist from 2012 onwards. The advised catch for the stock is reduced by the average proportion to give advice for the 7.e area.

This leads to the following advice for plaice in Division 7e:
Assuming the same proportion of the Division 7.e plaice stock is taken in Division 7.d as during 2003-2018, this will correspond to catches of plaice in Division 7.e of no more than 2456 tonnes. If discard rates do not change from the average of the last seven years (2012-2018), this implies landings of plaice in Division 7.e of no more than 1722 tonnes.

### 24.6 Biological reference points

The currently used reference points are the ones calculated at WKMSYREF4 (ICES, 2016a) and shown in the following table:

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}{ }^{*}$ | 2443 t | $\mathrm{B}_{\mathrm{pa}}$ | ICES <br> (2016a) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.238 | Eqsim run with segmented regression with breakpoint at $\mathrm{B}_{\text {loss }}$. $\mathrm{F}_{\text {MSY }}$ was taken as the peak of the median landings yield curve. | ICES <br> (2016a) |
| Precautionary approach | $\mathrm{Blim}_{\text {lim }}$ | 1745 t | $\mathrm{B}_{\text {loss }}$ | $\begin{aligned} & \text { ICES } \\ & (2016 \mathrm{a}) \end{aligned}$ |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 2443 t | $1.4 * \mathrm{~B}_{\text {lim }}$ | ICES (2016a) |
|  | $\mathrm{F}_{\text {lim }}$ | 0.88 | Based on segmented regression simulation of recruitment without error | ICES <br> (2016a) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.63 | Flim* $\exp \left(-1.645^{*} \sigma\right) ; \sigma=0.2$ | ICES <br> (2016a) |

* The value for MSY $B_{\text {trigger }}$ is not the value published in WKMSYREF4. The advice drafting group in 2017 and 2018 decided to base MSY $B_{\text {trigger }}$ on $B_{p a}$.

Reference points for 7.e plaice were calculated at WKMSYREF4 (ICES, 2016a) using the results from an XSA with parameters implemented at WGCSE 2015. In contrast to the WGCSE assessment 2015, absolute values from the XSA assessment were used instead of the relative values for the calculation of the values. ICES did not adopt these reference points initially due to the classification of the plaice 7.e as category 3 .

Instead MSY proxies were calculated at WKMS YPROXY2015 (ICES, 2016b) which are presented in the following table.

| Framework | Reference <br> point | Value | Technical basis | Source |
| :--- | :--- | :--- | :--- | :--- |
| MSY approach | MSY <br> proxy | 1910 triger | $\mathrm{F}_{\text {MSY }}$ (estimated by SPiCT from model parameters <br> using data from 1980-2014) | WKPROXY 2015 (ICES, <br> 2016b) |
|  | F $_{\text {MSY }}$ proxy | 0.56 | $0.5 \times B_{\text {MSY }}$ (estimated by SPiCT from model param- <br> eters using data from 1980-2014) | WKPROXY 2015 (ICES, <br> 2016b) |

These values have been used to assess the relative stock status at WGCSE 2016. At WGCSE 2017, these values have been rejected. Instead, the values from the WKMSYREF4 were restored and used in 2017, 2018 and 2019.

### 24.7 Exploratory assessment including discards

At the 2019 working group, an exploratory assessment including a full catch time-series, including discards, was trialled for the first time. Reliable discard data are available from 2012 onwards from InterCatch but not routinely used in the stock assessment.

For the years for which discard data are available, the raising was done inside InterCatch and for the strata for which no discard estimates were available, discards were estimated based on available discard information. Age allocations for discards were performed in InterCatch the same as for landings.

For the years prior to 2012, discards were estimated based on available discard information. Discard ratios at-age (for numbers) were calculated for 2012-2018 and a linear regression fitted, independently for each age used in the assessment. This was then extrapolated back in time to the beginning of the landings time-series in 1980. For the ages for which the linear regression did not indicate a decrease in the discards rate when going back in time, the average discard rate for 2012-2018 was used for all historical years prior to 2012. Discard weights-at-age where calculated as the mean of 2012-2018. Based on these discard numbers and weights-at-age, total discards were calculated, and the total catch corrected for discards. Total catch split into landings and discards, including the historical estimated discards are shown in Figure 23.24.
Based on the total catch estimates, an exploratory XSA assessment was conducted using identical XSA assessment settings as for the accepted landings only assessment. The results of this total catch assessment in comparison to the landings only assessment are shown in Figure 23.25. New reference points were calculated with EqSim with the same settings and assumptions as during WKMSYRef4 but using the results from the total catch assessment.

According to the total catch assessment and the revised reference points, the stock is lower than estimated by the landings only assessment and the recent decline is more pronounced. The fishing mortality is higher and above FMSY and FPa and below Flim.
It should be noted that this assessment is only exploratory and should not yet be used for providing advice. Further work and model validations are required.

The conclusion from this exploratory assessment run is that the stock is likely to be in less favourable condition as estimated by the accepted routine landings only assessment.

### 24.8 Management plans

There is no management plan in place for this stock.

### 24.9 Uncertainties and bias in assessment and forecast

A degree of uncertainty exists over the landings statistics for this stock given that mature plaice migrate between 7.d and 7.e during the spawning period. The current assessment applies a spawning migration correction that reallocates $15 \%$ of quarter 1 landings for the mature proportion of the catch from 7.d to 7.e. Consequently, the assessment results depend on the mixing rate assumption estimated from existing tagging data. Further work is required to examine stock structure and the mixing rate during the spawning period. Additional data are also needed to determine if the current mixing rateremains valid given the increased abundance of plaice stocks in the English Channel in recent years.
There is a heavy reliance on the age composition data derived from UK(E\&W) sampling. Around $25 \%$ of the landings for this stock are taken by countries that do not provide age-based data, and this situation is improved only slightly once the migration correction data from 7.d are added.
Discard data are only available for 2012-2018 and these data are mainly from the UK(E\&W). Historical discarding rates are highly uncertain but available discard data reported imply a significant increase between 2012 and 2015. Discards are not included in the assessment. The assessment contains a certain degree of uncertainty due to excluding discards and is likely to be overly optimistic. Fishing mortality is likely to be higher, and SSB lower than estimated by the current assessment. The decision to exclude discards in the assessment is based on the uncertainty in the available discards data and unknown discard survival rate of plaice.

### 24.10 Recommendations for next Benchmark

The perception of issues has not changed since last year's report. Therefore, the recommendations presented here, are identical to last year as they have not been addressed so far.

A benchmark assessment was developed for this stock at WKFLAT 2010 and an inter-benchmark meeting (IBPWCFlat2) subsequently convened in 2015 to revise the input data and update the XSA assessment settings. Nevertheless, any future benchmark meeting will need to consider the following issues.

Since 2017, ICES asked for the additional application of data-limited methods for category 3 stocks. This massively increased the workload for the stock coordinator and assessor but with little benefit for this stock. Upgrading this stock to category 1 is desirable and feasible within a reasonable timeframe.

The decisive reason for downgrading the stock to category 3 were unacceptable retrospective patterns in the XSA assessment. This has disappeared and a fully analytical assessment is possible. For doing so, the following issues need to be considered:

- A discard time-series should be developed and included into the assessment as discarding was substantial in recent years. The current assessment is based on landings only and therefore possibly fails to accurately model actual stock dynamics, particularly as the discard rate in recent years is variable.
- Discards including age compositions are now routinely estimated within InterCatch and exist for 2012-2017. Some UK discard data prior to 2012 exist, but have never been used. The discard time-series should be extended back in time, as it has been done for other plaice and similar stocks.
- Including discards in the assessment might require a re-parameterization of XSA settings and the exploration of alternative age-structured assessment models.
- Biological data such as natural mortality and maturity ogives are time invariant in the current assessment and borrowed from other plaice stocks (divisions 7.fg and 7.a). There
have been benchmarks for other plaice stocks and a similar approach could be pursuit for plaice in 7.e. The natural mortality used for plaicein7.e was originally borrowed from plaice in Division 7.a. The values for plaice in 7.a have been changed during the last benchmark but the original values are still used for plaice in 7.e.

Furthermore, the following points should be considered:

- Smoothing of stock and catch weights. The raw catch weights are corrected for migration from 7.d and then smoothed using a polynomial function of 2nd degree. Even though the fit seems to be quite reasonable different more appropriate methods should be evaluated.
- Abundance estimates derived from the UK FSP-7e and Q1SWBeam surveys included in the assessment are spatially restricted to the same areas as the commercial tuning fleets, and therefore little population abundance information exists along the French coast. Cpue estimates from additional research surveys in French coastal waters would improve the robustness of future assessment outputs.
- Investigate the addition of age composition information from the French and Belgian fleets. These fleets collectively account for about $16 \%$ of the total landings of this stock. In particular, inclusion of French data would add information on the stock dynamics on the French coast.


### 24.11 Management considerations

The stock unit (Division 7.e) does not correspond with the management unit (Divisions 7.d and 7.e), and this divisional mismatch hampers the effective management of plaice in the Western English Channel. However, some provision must be made to consider the effective management of adjacent plaice stocks given that components of the 7 .e stock are also taken during spawning period in 7.d. WKPLE 2015 revised the established migration correction, so that $15 \%$ of quarter 1 landings for the mature proportion of the catch are reallocated from 7.d to 7.e and the associated age composition is applied to plaice 7.e.

The total allowable catch (TAC) for the management area for 2016 has been doubled compared to 2015 but was reduced for 2017 and increased again slightly for 2018 and 2019.

Due to migration patterns, catches of this stock also occur in Division 7.d during the spawning period; therefore, to be consistent with the advised catch for the Division 7.e plaice stock, the actual catches of plaice in Division 7 .e should be lower than the advised catch for the stock. ICES has calculated the corresponding actual catches in Division 7.e, assuming that the proportion of Division 7.e stock catches taken in Division 7.d remains as in previous years (i.e. $10 \%$, the average of 2003-2018, taking the age structure of the population into account). As the mixing rate of the two plaice stocks is uncertain, this calculation provides only an approximation.

In accordance with the guidelines for category 3 stocks, a fully analytical assessment of the plaice 7.e stock including short-term forecast was not conducted at WGCSE 2018. Consequently, this year's category 3 assessment is indicative of trends only. Relative values presented for recruitment, spawning-stock biomass and fishing mortality estimates had similar temporal trends to absolute values presented at previous Working Groups.

The recent stock trends for plaice in 7.d are similar to plaice in 7.e.

### 24.12 References

EU. 2013. Regulation (EU) No 1380/2013 of the European Parliament and of the Council of 11 December 2013 on the Common Fisheries Policy, amending Council Regulations (EC) No 1954/2003 and (EC) No $1224 / 2009$ and repealing Council Regulations (EC) No 2371/2002 and (EC) No 639/2004 and Council Decision 2004/585/EC. OJ L 354, 28.12.2013, p. 22-61. http://data.europa.eu/eli/reg/2013/1380/oj
EU. 2017. Council Regulation. (EU) 2017/127 of 20 January 2017 fixing for 2017 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters. OJ L 24, 28.1.2017, p. 1-172. http://data.europa.eu/eli/reg/2017/127/oj

EU. 2018a. Commission Delegated Regulation (EU) 2018/2034 of 18 October 2018 establishing a discard plan for certain demersal fisheries in North-Western waters for the period 2019-2021. C/2018/6789.OJ L 327, 21.12.2018, p. 8-16. http://data.europa.eu/eli/reg del/2018/2034/oj

EU. 2018b. Council Regulation. (EU) 2018/120 of 23 January 2018 fixing for 2018 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters, and amending Regulation (EU) 2017/127. OJ L 27, 31.1.2018, p. 1168. http://data.europa.eu/eli/reg/2018/120/oj

EU. 2019. Council Regulation (EU) 2019/124 of 30 January 2019 fixing for 2019 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters. ST/15733/2018/INIT. OJ L 29, 31.1.2019, p. 1-166. http://data.europa.eu/eli/reg/2019/124/oj

ICES. 2014. Report of the Workshop on the Development of Quantitative Assessment Methodologies based on LIFE-history traits, exploitation characteristics, and other relevant parameters for data-limited stocks (WKLIFE IV), 27-31 October 2014, Lisbon, Portugal. ICES CM 2014/ACOM:54. 223 pp.
ICES. 2015e. Report of the Fifth Workshop on the Development of Quantitative Assessment Methodologies based on Life-history Traits, Exploitation Characteristics and other Relevant Parameters for Data-limited Stocks (WKLIFE V), 5-9 October 2015, Lisbon, Portugal. ICES CM 2015/ACOM:56. 157 pp.

ICES. 2016a. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMS YREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

ICES. 2016b. Report of the Workshop to consider MSY proxies for stocks in ICES category 3 and 4 stocks in Western Waters (WKProxy), 3-6 November 2015, ICES Headquarters, Copenhagen. ICES CM 2015/ACOM:61. 183 pp.

Pedersen, M. W., and Berg, C. W. 2017. A stochastic surplus production model in continuous time. Fish and Fisheries, 18(2), 226-243.

### 24.13 Tables

Table 23.1. Plaice in 7.e. Nominal landings (t) in Division 7e, as used by the Working Group.

|  | Landi |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \frac{1}{\pi} \\ & \stackrel{\text { ® }}{0} \end{aligned}$ | $\begin{aligned} & \frac{E}{5} \\ & \frac{0}{60} \\ & \infty \end{aligned}$ |  | $\begin{aligned} & \text { U } \\ & \text { U } \\ & \text { 뀬 } \end{aligned}$ |  | $\begin{aligned} & \stackrel{\sim}{む} \\ & \stackrel{5}{ \pm} \end{aligned}$ | $\begin{aligned} & \text { O} \\ & \text { N } \\ & \text { O} \\ & \text { N } \\ & \text { N } \\ & \stackrel{0}{0} \end{aligned}$ |  | $\begin{aligned} & \stackrel{*}{\sqrt[\pi]{0}} \\ & \stackrel{0}{\circ} \end{aligned}$ |  |  | $\begin{aligned} & * \\ & \stackrel{*}{*} \\ & \stackrel{n}{n} \\ & \stackrel{0}{0} \\ & \stackrel{H}{0} \\ & \ddot{0} \end{aligned}$ |
| 1976 | 5 | - | 323 | 312 | - | 640 | - | 640 | - | 640 |  |
| 1977 | 3 | - | 336 | 363 | - | 702 | - | 702 | - | 702 |  |
| 1978 | 3 | - | 314 | 467 | - | 784 | - | 784 | - | 784 |  |
| 1979 | 2 | - | 458 | 515 | - | 975 | 2 | 977 | - | 977 |  |
| 1980 | 23 | - | 325 | 609 | 9 | 966 | 113 | 1079 | 99 | 1178 |  |
| 1981 | 27 | - | 537 | 953 | - | 1517 | -16 | 1501 | 175 | 1676 |  |
| 1982 | 81 | - | 363 | 1109 | - | 1553 | 135 | 1688 | 190 | 1878 |  |
| 1983 | 20 | - | 371 | 1195 | - | 1586 | -91 | 1495 | 219 | 1714 |  |
| 1984 | 24 | - | 278 | 1144 | - | 1446 | 101 | 1547 | 211 | 1758 |  |
| 1985 | 39 | - | 197 | 1122 | - | 1358 | 83 | 1441 | 236 | 1677 |  |
| 1986 | 26 | - | 276 | 1389 | - | 1691 | 119 | 1810 | 268 | 2078 |  |
| 1987 | 68 | - | 435 | 1419 | - | 1922 | 36 | 1958 | 314 | 2272 |  |
| 1988 | 90 | - | 584 | 1654 | - | 2328 | 130 | 2458 | 377 | 2835 |  |
| 1989 | 89 | - | 448 | 1712 | - | 2249 | 109 | 2358 | 384 | 2742 |  |
| 1990 | 82 | - | N/A | 1891 | 2 | 1977 | 616 | 2593 | 392 | 2985 |  |
| 1991 | 57 | - | 251 | 1326 | - | 1634 | 214 | 1848 | 335 | 2183 |  |
| 1992 | 25 | - | 419 | 1110 | 14 | 1568 | 56 | 1624 | 258 | 1882 |  |
| 1993 | 56 | - | 284 | 1080 | 24 | 1444 | -27 | 1417 | 197 | 1614 |  |
| 1994 | 10 | - | 277 | 998 | - | 1285 | -129 | 1156 | 248 | 1404 |  |
| 1995 | 13 | - | 288 | 857 | - | 1158 | -127 | 1031 | 216 | 1247 |  |
| 1996 | 4 | - | 279 | 855 | - | 1138 | -94 | 1044 | 222 | 1266 |  |
| 1997 | 6 | - | 329 | 1038 | 1 | 1374 | -51 | 1323 | 260 | 1583 |  |
| 1998 | 22 | - | 327 | 892 | 1 | 1242 | -111 | 1131 | 215 | 1346 |  |
| 1999 | 12 | - | 194 | 947 | - | 1153 | 146 | 1299 | 244 | 1543 |  |


|  | Landings |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\stackrel{\grave{1}}{\stackrel{1}{\sim}}$ | $\begin{aligned} & E \\ & \frac{\varepsilon}{\bar{W}} \\ & \frac{0}{0} \\ & \infty \end{aligned}$ |  | $\begin{aligned} & \stackrel{U}{C} \\ & \stackrel{\pi}{4} \end{aligned}$ |  | $\begin{aligned} & \stackrel{N}{0} \\ & \stackrel{5}{5} \end{aligned}$ | $\begin{aligned} & \ddot{0} \\ & \stackrel{t}{0} \\ & \stackrel{0}{2} \\ & \frac{2}{7} \\ & \stackrel{0}{0} \end{aligned}$ |  | $\begin{aligned} & \frac{*}{\sqrt[\pi]{0}} \\ & \stackrel{0}{\circ} \end{aligned}$ |  |  |  |
| 2000 | 4 | － | 360 | 926 | ＋ | 1290 | －9 | 1281 | 345 | 1625 |  |
| 2001 | 12 | － | 303 | 797 | － | 1112 | －6 | 1106 | 204 | 1310 |  |
| 2002 | 27 | － | 242 | 978 | ＋ | 1247 | 10 | 1257 | 215 | 1472 |  |
| 2003 | 39 | － | 216 | 985 | － | 1240 | 37 | 1277 | 110 | 1387 |  |
| 2004 | 46 | － | 184 | 912 | － | 1142 | 70 | 1212 | 126 | 1337 |  |
| 2005 | 48 | － | 198 | 887 | － | 1133 | 70 | 1203 | 117 | 1319 |  |
| 2006 | 52 | － | 223 | 964 | － | 1239 | 74 | 1313 | 97 | 1411 |  |
| 2007 | 84 | － | 202 | 680 | － | 966 | 37 | 1003 | 143 | 1146 |  |
| 2008 | 66 | － | 148 | 676 | － | 890 | 86 | 976 | 135 | 1112 |  |
| 2009 | 53 | 2 | 191 | 729 | － | 975 | －52 | 923 | 101 | 1024 |  |
| 2010 | 51 | 2 | 227 | 843 | － | 1123 | －31 | 1092 | 116 | 1208 |  |
| 2011 | 141 | 3 | 274 | 936 | － | 1354 | －20 | 1334 | 83 | 1417 |  |
| 2012 | 134 | 2 | 224 | 1003 | － | 1363 | 3 | 1366 | 126 | 1492 | 448 |
| 2013 | 97 | 1 | 221 | 1041 | － | 1360 | －9 | 1351 | 121 | 1472 | 351 |
| 2014 | 41 | 0 | 323 | 976 | － | 1340 | 1 | 1341 | 149 | 1490 | 1133 |
| 2015 | 111 | 1 | 224 | 912 | 1 | 1249 | 3 | 1246 | 178 | 1424 | 1276 |
| 2016 | 145 | ＜ 1 | 204 | 1430 | － | 1780 | －1 | 1777 | 235 | 2013 | 618 |
| 2017 | 151 | ＜1 | 153 | 1602 | 1 | 1911 | 4 | 1915 | 213 | 2128 | 821 |
| 2018＾ | 142 | 3 | 118 | 1373 | － | 1640 | 4 | 1644 | 236 | 1880 | 624 |

＊Estimated by the working group．
${ }^{* *}$ Migration correction（ $15 \%$ of the mature population caught in Quarter 1 in Division 7．d）added to stock．
＊＊＊Discard estimated by the working group，including discards from the migration correction．
${ }^{\wedge}$ Preliminary．

Table 23.2. Plaice in 7.e. Landings numbers-at-age.

| year/age | Numbers-at-age [thousands] |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | TOTAL NUM |
| 1980 | 754 | 758 | 244 | 226 | 62 | 63 | 22 | 13 | 137 | 2279 |
| 1981 | 667 | 2068 | 555 | 118 | 101 | 20 | 46 | 18 | 94 | 3688 |
| 1982 | 279 | 1928 | 1371 | 257 | 87 | 82 | 16 | 28 | 121 | 4168 |
| 1983 | 720 | 799 | 1613 | 586 | 101 | 40 | 47 | 2 | 99 | 4009 |
| 1984 | 928 | 1650 | 659 | 518 | 191 | 90 | 28 | 33 | 50 | 4146 |
| 1985 | 596 | 1424 | 1326 | 154 | 248 | 140 | 27 | 15 | 51 | 3980 |
| 1986 | 914 | 2326 | 908 | 478 | 110 | 127 | 66 | 28 | 61 | 5018 |
| 1987 | 1063 | 2083 | 1355 | 648 | 228 | 86 | 49 | 44 | 51 | 5608 |
| 1988 | 1817 | 4627 | 1087 | 456 | 149 | 112 | 38 | 24 | 52 | 8362 |
| 1989 | 269 | 2748 | 2873 | 825 | 268 | 118 | 94 | 31 | 100 | 7326 |
| 1990 | 331 | 3151 | 2668 | 1198 | 263 | 133 | 76 | 56 | 71 | 7946 |
| 1991 | 557 | 1192 | 1876 | 956 | 510 | 103 | 43 | 33 | 51 | 5320 |
| 1992 | 699 | 1299 | 734 | 646 | 441 | 258 | 69 | 32 | 49 | 4227 |
| 1993 | 670 | 1377 | 631 | 262 | 267 | 216 | 165 | 39 | 85 | 3712 |
| 1994 | 326 | 1503 | 831 | 250 | 106 | 116 | 78 | 84 | 63 | 3357 |
| 1995 | 322 | 732 | 943 | 263 | 118 | 56 | 79 | 68 | 88 | 2667 |
| 1996 | 1050 | 668 | 379 | 382 | 122 | 59 | 38 | 47 | 105 | 2848 |
| 1997 | 861 | 2228 | 435 | 177 | 147 | 75 | 31 | 17 | 99 | 4070 |
| 1998 | 536 | 1482 | 1107 | 155 | 64 | 60 | 22 | 21 | 61 | 3507 |
| 1999 | 650 | 2135 | 1124 | 407 | 92 | 37 | 39 | 17 | 45 | 4546 |
| 2000 | 351 | 1157 | 2037 | 496 | 181 | 38 | 14 | 22 | 52 | 4348 |
| 2001 | 469 | 785 | 788 | 950 | 145 | 79 | 19 | 11 | 37 | 3283 |
| 2002 | 1017 | 1190 | 460 | 394 | 456 | 106 | 42 | 12 | 40 | 3718 |
| 2003 | 886 | 964 | 532 | 182 | 166 | 236 | 58 | 45 | 38 | 3107 |
| 2004 | 471 | 1364 | 566 | 338 | 107 | 74 | 109 | 51 | 38 | 3119 |
| 2005 | 796 | 880 | 775 | 277 | 146 | 50 | 49 | 58 | 48 | 3080 |
| 2006 | 995 | 1358 | 517 | 379 | 115 | 61 | 27 | 18 | 53 | 3523 |


| year/age | Numbers-at-age [thousands] |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | TOTAL NUM |
| 2007 | 393 | 1077 | 699 | 287 | 199 | 72 | 31 | 10 | 50 | 2819 |
| 2008 | 919 | 703 | 570 | 259 | 112 | 87 | 32 | 15 | 29 | 2727 |
| 2009 | 647 | 1255 | 297 | 151 | 79 | 32 | 21 | 7 | 17 | 2505 |
| 2010 | 759 | 974 | 758 | 215 | 114 | 47 | 16 | 18 | 23 | 2924 |
| 2011 | 1132 | 1441 | 725 | 255 | 75 | 50 | 27 | 12 | 18 | 3735 |
| 2012 | 204 | 1561 | 1066 | 373 | 253 | 101 | 51 | 21 | 35 | 3664 |
| 2013 | 137 | 1075 | 1377 | 510 | 200 | 149 | 45 | 49 | 36 | 3579 |
| 2014 | 135 | 636 | 1407 | 845 | 356 | 135 | 70 | 54 | 35 | 3673 |
| 2015 | 90 | 392 | 642 | 924 | 553 | 234 | 61 | 50 | 35 | 2982 |
| 2016 | 61 | 888 | 1116 | 828 | 897 | 426 | 155 | 64 | 55 | 4490 |
| 2017 | 88 | 460 | 1619 | 1148 | 646 | 468 | 220 | 133 | 134 | 4917 |
| 2018 | 72 | 392 | 913 | 1307 | 713 | 339 | 226 | 122 | 114 | 4199 |

Table 23.3. Plaice in 7.e. Landings weights-at-age.

| year/age | Weights-at-age [kg] |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1980 | 0.329 | 0.435 | 0.538 | 0.640 | 0.741 | 0.840 | 0.939 | 1.035 | 1.392 |
| 1981 | 0.273 | 0.400 | 0.526 | 0.647 | 0.767 | 0.883 | 0.997 | 1.108 | 1.448 |
| 1982 | 0.302 | 0.391 | 0.474 | 0.548 | 0.617 | 0.678 | 0.732 | 0.780 | 0.890 |
| 1983 | 0.224 | 0.338 | 0.446 | 0.547 | 0.642 | 0.730 | 0.812 | 0.888 | 1.085 |
| 1984 | 0.254 | 0.356 | 0.461 | 0.570 | 0.682 | 0.797 | 0.914 | 1.034 | 1.510 |
| 1985 | 0.222 | 0.337 | 0.450 | 0.561 | 0.669 | 0.775 | 0.878 | 0.979 | 1.341 |
| 1986 | 0.260 | 0.353 | 0.450 | 0.551 | 0.655 | 0.764 | 0.877 | 0.994 | 1.490 |
| 1987 | 0.287 | 0.347 | 0.418 | 0.503 | 0.599 | 0.710 | 0.833 | 0.968 | 1.387 |
| 1988 | 0.225 | 0.310 | 0.407 | 0.515 | 0.634 | 0.765 | 0.906 | 1.059 | 1.398 |
| 1989 | 0.224 | 0.293 | 0.370 | 0.454 | 0.547 | 0.647 | 0.756 | 0.872 | 1.167 |
| 1990 | 0.270 | 0.315 | 0.371 | 0.437 | 0.514 | 0.602 | 0.700 | 0.809 | 1.081 |
| 1991 | 0.252 | 0.316 | 0.389 | 0.473 | 0.566 | 0.670 | 0.784 | 0.908 | 1.246 |
| 1992 | 0.286 | 0.345 | 0.417 | 0.503 | 0.601 | 0.713 | 0.838 | 0.976 | 1.330 |
| 1993 | 0.263 | 0.338 | 0.418 | 0.503 | 0.596 | 0.694 | 0.798 | 0.907 | 1.194 |
| 1994 | 0.266 | 0.336 | 0.412 | 0.494 | 0.582 | 0.676 | 0.775 | 0.879 | 1.136 |
| 1995 | 0.282 | 0.362 | 0.445 | 0.531 | 0.619 | 0.709 | 0.803 | 0.899 | 1.083 |
| 1996 | 0.268 | 0.371 | 0.474 | 0.577 | 0.681 | 0.786 | 0.891 | 0.997 | 1.216 |
| 1997 | 0.272 | 0.345 | 0.427 | 0.514 | 0.608 | 0.709 | 0.816 | 0.931 | 1.196 |
| 1998 | 0.190 | 0.313 | 0.435 | 0.556 | 0.674 | 0.793 | 0.911 | 1.028 | 1.339 |
| 1999 | 0.206 | 0.295 | 0.382 | 0.466 | 0.548 | 0.628 | 0.706 | 0.781 | 1.006 |
| 2000 | 0.206 | 0.293 | 0.380 | 0.468 | 0.555 | 0.642 | 0.729 | 0.817 | 1.066 |
| 2001 | 0.218 | 0.301 | 0.388 | 0.480 | 0.576 | 0.677 | 0.782 | 0.891 | 1.268 |
| 2002 | 0.256 | 0.331 | 0.410 | 0.496 | 0.588 | 0.686 | 0.788 | 0.895 | 1.208 |
| 2003 | 0.266 | 0.371 | 0.475 | 0.577 | 0.675 | 0.772 | 0.866 | 0.959 | 1.273 |
| 2004 | 0.300 | 0.361 | 0.429 | 0.505 | 0.588 | 0.679 | 0.778 | 0.883 | 1.203 |
| 2005 | 0.293 | 0.366 | 0.445 | 0.528 | 0.616 | 0.709 | 0.806 | 0.908 | 1.134 |
| 2006 | 0.296 | 0.361 | 0.433 | 0.512 | 0.600 | 0.694 | 0.795 | 0.904 | 1.121 |


| year/age | Weights-at-age [kg] |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 2007 | 0.255 | 0.333 | 0.415 | 0.499 | 0.586 | 0.677 | 0.770 | 0.868 | 1.105 |
| 2008 | 0.281 | 0.357 | 0.441 | 0.531 | 0.627 | 0.729 | 0.838 | 0.954 | 1.308 |
| 2009 | 0.242 | 0.379 | 0.513 | 0.644 | 0.771 | 0.894 | 1.013 | 1.128 | 1.383 |
| 2010 | 0.274 | 0.364 | 0.460 | 0.562 | 0.668 | 0.779 | 0.895 | 1.016 | 1.285 |
| 2011 | 0.241 | 0.351 | 0.463 | 0.577 | 0.693 | 0.811 | 0.931 | 1.052 | 1.376 |
| 2012 | 0.207 | 0.310 | 0.413 | 0.515 | 0.618 | 0.721 | 0.824 | 0.927 | 1.239 |
| 2013 | 0.268 | 0.318 | 0.382 | 0.458 | 0.548 | 0.650 | 0.766 | 0.894 | 1.355 |
| 2014 | 0.207 | 0.280 | 0.358 | 0.441 | 0.528 | 0.619 | 0.714 | 0.814 | 1.164 |
| 2015 | 0.244 | 0.306 | 0.380 | 0.466 | 0.563 | 0.672 | 0.792 | 0.923 | 1.251 |
| 2016 | 0.279 | 0.325 | 0.379 | 0.441 | 0.512 | 0.591 | 0.677 | 0.773 | 1.001 |
| 2017 | 0.270 | 0.310 | 0.357 | 0.411 | 0.472 | 0.540 | 0.616 | 0.697 | 0.926 |
| 2018 | 0.219 | 0.262 | 0.322 | 0.397 | 0.487 | 0.594 | 0.716 | 0.855 | 1.177 |

Table 23.4. Plaice in 7.e. Stock weights-at-age.

| year/age | Stock weights-at-age [kg] |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1980 | 0.275 | 0.381 | 0.485 | 0.587 | 0.688 | 0.788 | 0.886 | 0.983 | 1.342 |
| 1981 | 0.207 | 0.336 | 0.462 | 0.585 | 0.705 | 0.823 | 0.937 | 1.049 | 1.393 |
| 1982 | 0.253 | 0.345 | 0.430 | 0.508 | 0.579 | 0.643 | 0.701 | 0.751 | 0.874 |
| 1983 | 0.164 | 0.282 | 0.393 | 0.497 | 0.595 | 0.687 | 0.772 | 0.851 | 1.059 |
| 1984 | 0.202 | 0.302 | 0.405 | 0.512 | 0.621 | 0.733 | 0.849 | 0.967 | 1.433 |
| 1985 | 0.163 | 0.280 | 0.394 | 0.506 | 0.615 | 0.722 | 0.827 | 0.929 | 1.295 |
| 1986 | 0.215 | 0.306 | 0.401 | 0.500 | 0.603 | 0.709 | 0.820 | 0.935 | 1.422 |
| 1987 | 0.261 | 0.313 | 0.378 | 0.455 | 0.545 | 0.648 | 0.764 | 0.892 | 1.292 |
| 1988 | 0.186 | 0.266 | 0.357 | 0.460 | 0.573 | 0.698 | 0.833 | 0.980 | 1.309 |
| 1989 | 0.193 | 0.258 | 0.330 | 0.411 | 0.500 | 0.596 | 0.701 | 0.813 | 1.098 |
| 1990 | 0.250 | 0.290 | 0.340 | 0.401 | 0.472 | 0.554 | 0.647 | 0.750 | 1.009 |
| 1991 | 0.224 | 0.282 | 0.350 | 0.428 | 0.516 | 0.615 | 0.723 | 0.842 | 1.167 |
| 1992 | 0.259 | 0.310 | 0.375 | 0.453 | 0.544 | 0.648 | 0.765 | 0.895 | 1.231 |
| 1993 | 0.227 | 0.298 | 0.375 | 0.458 | 0.547 | 0.641 | 0.742 | 0.848 | 1.126 |
| 1994 | 0.230 | 0.297 | 0.369 | 0.447 | 0.531 | 0.620 | 0.715 | 0.816 | 1.063 |
| 1995 | 0.243 | 0.322 | 0.403 | 0.487 | 0.573 | 0.663 | 0.755 | 0.850 | 1.031 |
| 1996 | 0.217 | 0.319 | 0.421 | 0.524 | 0.628 | 0.732 | 0.837 | 0.943 | 1.160 |
| 1997 | 0.237 | 0.308 | 0.385 | 0.469 | 0.559 | 0.657 | 0.761 | 0.872 | 1.129 |
| 1998 | 0.128 | 0.251 | 0.374 | 0.495 | 0.616 | 0.735 | 0.853 | 0.971 | 1.283 |
| 1999 | 0.160 | 0.250 | 0.339 | 0.424 | 0.508 | 0.589 | 0.667 | 0.743 | 0.972 |
| 2000 | 0.162 | 0.248 | 0.335 | 0.422 | 0.509 | 0.596 | 0.683 | 0.771 | 1.019 |
| 2001 | 0.178 | 0.259 | 0.344 | 0.434 | 0.528 | 0.626 | 0.729 | 0.836 | 1.205 |
| 2002 | 0.215 | 0.285 | 0.361 | 0.443 | 0.529 | 0.621 | 0.719 | 0.822 | 1.119 |
| 2003 | 0.211 | 0.318 | 0.422 | 0.524 | 0.624 | 0.722 | 0.817 | 0.911 | 1.227 |
| 2004 | 0.272 | 0.329 | 0.393 | 0.464 | 0.544 | 0.630 | 0.725 | 0.827 | 1.136 |
| 2005 | 0.257 | 0.328 | 0.404 | 0.484 | 0.569 | 0.659 | 0.754 | 0.853 | 1.074 |
| 2006 | 0.265 | 0.326 | 0.395 | 0.471 | 0.554 | 0.644 | 0.741 | 0.846 | 1.057 |


| year/age | Stock weights-at-age [kg] |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 2007 | 0.217 | 0.294 | 0.374 | 0.457 | 0.542 | 0.631 | 0.723 | 0.818 | 1.052 |
| 2008 | 0.245 | 0.318 | 0.398 | 0.484 | 0.577 | 0.676 | 0.782 | 0.894 | 1.238 |
| 2009 | 0.171 | 0.311 | 0.447 | 0.579 | 0.707 | 0.832 | 0.953 | 1.070 | 1.329 |
| 2010 | 0.229 | 0.318 | 0.411 | 0.509 | 0.612 | 0.720 | 0.834 | 0.952 | 1.215 |
| 2011 | 0.186 | 0.295 | 0.407 | 0.520 | 0.635 | 0.752 | 0.870 | 0.991 | 1.313 |
| 2012 | 0.156 | 0.259 | 0.361 | 0.464 | 0.567 | 0.670 | 0.773 | 0.876 | 1.187 |
| 2013 | 0.247 | 0.291 | 0.348 | 0.418 | 0.501 | 0.597 | 0.706 | 0.828 | 1.270 |
| 2014 | 0.172 | 0.243 | 0.319 | 0.399 | 0.484 | 0.573 | 0.666 | 0.763 | 1.119 |
| 2015 | 0.217 | 0.274 | 0.342 | 0.422 | 0.513 | 0.616 | 0.730 | 0.856 | 1.181 |
| 2016 | 0.259 | 0.301 | 0.350 | 0.409 | 0.475 | 0.550 | 0.633 | 0.725 | 0.948 |
| 2017 | 0.252 | 0.289 | 0.333 | 0.383 | 0.441 | 0.505 | 0.577 | 0.655 | 0.881 |
| 2018 | 0.203 | 0.239 | 0.29 | 0.357 | 0.44 | 0.539 | 0.653 | 0.784 | 1.097 |

Table 23.5. Plaice in 7.e. Tuning fleet data available. Not all years and ages as shown here are used in the assessment.


| 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 |  |
|  | 0 | 0.2730774994 | 0.765289223 | 0.9806866884 | 0.6563497784 |  |
|  | 0.1850327432 | 0.1112903326 | 0.0873483927 | 0.0302633863 | 0.0263229229 |  |
|  | 0.0030350849 | 0.0017983016 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| Q1SWBeam |  |  |  |  |  |  |
| 20062018 |  |  |  |  |  |  |
| 1100.25 |  |  |  |  |  |  |
| 127 |  |  |  |  |  |  |
| 1 | 1.46029 | 31.1894 | 24.244 | 19.115 | 5.3835 |  |
|  | 2.6963 | 0.1513 | 0.11942 | 0.2388 | 0.56317 | 0 |
|  | 0.34656 | 0.1976 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0.86782 | 14.7809 | 34.368 | 28.319 | 4.9883 |  |
|  | 5.5958 | 1.9261 | 4.75535 | 0.2503 | 3.992 |  |
|  | 0.2503 | 2.29854 | 0.99913 | 0.2503 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.95099 | 33.5532 | 17.429 | 9.116 | 5.4635 |  |
|  | 0.9659 | 1.5218 | 2.21499 | 1.979 | 0 |  |
|  | 0.87797 | 0.12102 | 0.18772 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 1.2131 | 45.2746 | 46.545 | 15.717 | 10.7145 |  |
|  | 3.0017 | 4.1608 | 0.32375 | 0.2043 | 0.32375 |  |
|  | 0.32375 | 0 | 1.78752 | 0 | 1.23777 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 5.27731 | 0 | 0 | 0 | 0 |  |
| 1 | 0.97592 | 45.0547 | 39.746 | 27.094 | 4.3481 |  |
|  | 1.8618 | 2.7469 | 0.76424 | 0.3754 | 0 | 0 |
|  | 0.18772 | 0 | 2.29257 | 0 | 0 | 0 |
|  | 0 | 0.11942 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 1.80958 | 53.1512 | 75.65 | 27.139 | 7.0529 |  |
|  | 6.2411 | 2.9599 | 0.46795 | 0.5277 | 0.11377 |  |
|  | 0.35757 | 0 | 0.11919 | 0.11919 | 0.11919 |  |
|  | 0.64983 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 9.1228 | 59.258 | 30.977 | 14.8202 |  |
|  | 5.2353 | 7.4435 | 0.48139 | 3.1713 | 0 | 0 |
|  | 0.18772 | 0 | 0 | 0.14385 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.11942 | 0 |  |  |
| 1 | 0.30036 | 18.0403 | 91.824 | 65.429 | 12.689 |  |
|  | 3.9641 | 2.5307 | 2.00951 | 0.8034 | 0 | 0 |
|  | 0 | 0.12571 | 0 | 0 | 0 |  |
|  | 0.11374 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 1.01423 | 65.9025 | 148.705 | 178.597 | $63.2579$ |  |
|  | 10.6805 | 1.3356 | 2.33955 | 0.9387 | $0.48829$ |  |
|  | 0.28101 | 0.15884 | 0.1706 | 0 | 0 | 0 |
|  | 0.15884 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 36.3433 | 46.731 | 27.17 | 40.4109 |  |
|  | 30.2577 | 4.3911 | 5.31769 | 0.9476 | 2.08315 | 0 |
|  | 0.99022 | 0 | 0.18772 | 0 | 0.1976 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0.22085 | 20.8393 | 190.215 | 56.534 | 34.9048 |  |
|  | 37.2433 | 26.557 | 7.92202 | 11.0153 | 1.75709 |  |
|  | 6.95574 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 14.2514 | 68.17 | 151.523 | 32.551 |  |
|  | 31.4779 | 14.6612 | 2.65757 | 5.4495 | 0.37686 |  |
|  | 5.39347 | 0 | 0.30715 | 0 | 0.18772 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 11.2905 | 37.574 | 27.32 | 45.9335 |  |
|  | 22.3257 | 10.4298 | 9.19479 | 9.1063 | 1.1739 |  |
|  | 1.07603 | 0.11942 | 0 | 0.11942 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| FSP-7e-biomass |  |  |  |  |  |  |
| 20032018 |  |  |  |  |  |  |
| $110.750 .80$ |  |  |  |  |  |  |


| 1 | 0 | 0.017016728 | 0.068867092 | 0.096605209 | 0.026719823 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.030511966 | 0.039428295 | 0.023136956 | 0.02121698 | 0.003105212 |  |
|  | 0.001649358 | 0.00101558 | 0.000129111 | 6.16112e-05 | 0 |  |
|  | 6.16112e-05 | 0 | 6.16112e-05 | 0.000412429 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0.009253684 | 0.14719821 | 0.077734516 | 0.142615388 |  |
|  | 0.019093729 | 0.013918673 | 0.045087957 | 0.009428396 | 0.004482924 |  |
|  | 0.001121228 | 0.000215559 | 0.000182454 | 0 | 0.000304188 | 0 |
|  | 0 | 0 | 0.000240839 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.015621844 | 0.073065687 | 0.101896379 | 0.05382426 |  |
|  | 0.027075454 | 0.010259167 | 0.017859677 | 0.034554905 | 0.007686843 |  |
|  | 0.006065153 | 0 | 0 | 0 | 0.001218691 |  |
|  | 0.000962531 | 0.001353022 | 0 | 0.001294305 | 0.002771882 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0.026371445 | 0.068631235 | 0.085981713 | 0.059904288 |  |
|  | 0.029057905 | 0.015892962 | 0.00406511 | 0.006358017 | 0.015281921 |  |
|  | 0.002559487 | 0.003623002 | 0 | 0 | 0.000347422 | 0 |
|  | 0 | 0.001102939 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.004778692 | 0.047412485 | 0.063290284 | 0.03830767 |  |
|  | 0.020106321 | 0.007997441 | 0.003733733 | 0.005285688 | 0.005281424 |  |
|  | 0.002276312 | 0.004246414 | 0.001312233 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.02303336 | 0.042696327 | 0.06099742 | 0.022717433 |  |
|  | 0.016696799 | 0.014316446 | 0.006020215 | 0.001612963 | 0.000686571 |  |
|  | 0.000219825 | 0.000258466 | $9.83678 \mathrm{e}-05$ | 3.66375e-05 | $6.17303 \mathrm{e}-05$ |  |
|  | 3.66375e-05 | 0 | $6.17303 \mathrm{e}-05$ | 6.17303e-05 | $6.17303 \mathrm{e}-05$ | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0.021783708 | 0.089427434 | 0.071486272 | 0.040609123 |  |
|  | 0.031037596 | 0.018685442 | 0.00724396 | 0.005443947 | 0.00255305 | 0 |
|  | 0.003877076 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0 | 0.031128145 | 0.108246974 | 0.176676563 | 0.032028353 |  |
|  | 0.028385365 | 0.011556535 | 0.006137064 | 0.002962591 | 0.00716298 |  |
|  | 0.002042332 | 0.003041061 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.039626333 | 0.124140706 | 0.126023141 | 0.065743914 |  |
|  | 0.006395157 | 0.01993035 | 0.010885693 | 0.002822451 | 0.002837335 |  |
|  | 0.002214251 | 0.000167923 | 0.000106658 | 0.004218814 | 0 |  |
|  | 0.001542561 | 0 | $6.12651 \mathrm{e}-05$ | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0.01139495 | 0.180226935 | 0.207399701 | 0.07492458 |  |
|  | 0.080259232 | 0.013262699 | 0.009394429 | 0.011571324 | 0.002465811 |  |
|  | 0.005859429 | 0.004732132 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.006907976 | 0.180758085 | 0.410553109 | 0.202774124 |  |
|  | 0.108655459 | 0.054254917 | 0.020215075 | 0.010110847 | 0.003552844 | 0 |
|  | 0.001455044 | 0.001455044 | 0 | 0 | 0 | 0 |
|  | 0.003290293 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0 | 0.041444759 | 0.205551436 | 0.631560929 | 0.333907762 |  |
|  | 0.261575403 | 0.079435747 | 0.025108143 | 0.029091589 | 0 |  |
|  | 0.004970024 | 0.001837137 | 0 | 0 | 0.000926584 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.025917135 | 0.177782812 | 0.309264643 | 0.41997488 |  |
|  | 0.249988758 | 0.076811998 | 0.012487193 | 0.03663801 | 0.001327924 |  |
|  | 0.012810427 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.012422358 | 0.194476531 | 0.290658509 | 0.183274879 |  |
|  | 0.180552834 | 0.125276737 | 0.013959313 | 0.045946967 | 0 |  |
|  | 0.00886356 | 0 | 0.001943963 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.008907602 | 0.07598728 | 0.484763131 | 0.149260328 |  |
|  | 0.207245338 | 0.155649866 | 0.03381067 | 0.09188589 | 0.012446502 |  |
|  | 0.019714288 | 0.010609453 | 0.005307667 | 0 | 0.003212512 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.00926128 | 0.095714846 | 0.287166203 | 0.277926644 |  |
|  | 0.105983836 | 0.068842772 | 0.069157464 | 0.022306102 | 0.023142302 |  |
|  | 0.003509849 | 0.003526564 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |

Table 23.6. Plaice in 7.e. Fishing mortality-at-age. The values in the table are rounded to three digits.

| year/age | Fishing mortality-at-age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | $F(3-6)$ |
| 1980 | 0.12 | 0.419 | 0.457 | 0.423 | 0.766 | 0.407 | 0.341 | 0.507 | 0.507 | 0.516 |
| 1981 | 0.107 | 0.503 | 0.562 | 0.378 | 0.309 | 0.553 | 0.54 | 0.469 | 0.469 | 0.438 |
| 1982 | 0.104 | 0.461 | 0.67 | 0.502 | 0.481 | 0.401 | 1.073 | 0.655 | 0.655 | 0.528 |
| 1983 | 0.128 | 0.436 | 0.803 | 0.616 | 0.342 | 0.392 | 0.389 | 0.375 | 0.375 | 0.549 |
| 1984 | 0.187 | 0.433 | 0.71 | 0.591 | 0.375 | 0.525 | 0.469 | 0.458 | 0.458 | 0.527 |
| 1985 | 0.095 | 0.438 | 0.676 | 0.318 | 0.571 | 0.474 | 0.261 | 0.437 | 0.437 | 0.501 |
| 1986 | 0.144 | 0.58 | 0.504 | 0.498 | 0.358 | 0.585 | 0.39 | 0.446 | 0.446 | 0.485 |
| 1987 | 0.08 | 0.508 | 0.727 | 0.748 | 0.427 | 0.477 | 0.425 | 0.445 | 0.445 | 0.602 |
| 1988 | 0.174 | 0.523 | 0.493 | 0.52 | 0.341 | 0.348 | 0.361 | 0.351 | 0.351 | 0.469 |
| 1989 | 0.033 | 0.392 | 0.656 | 0.789 | 0.602 | 0.452 | 0.501 | 0.521 | 0.521 | 0.61 |
| 1990 | 0.101 | 0.593 | 0.746 | 0.572 | 0.566 | 0.616 | 0.531 | 0.574 | 0.574 | 0.619 |
| 1991 | 0.164 | 0.568 | 0.784 | 0.594 | 0.463 | 0.409 | 0.376 | 0.418 | 0.418 | 0.602 |
| 1992 | 0.184 | 0.631 | 0.757 | 0.62 | 0.548 | 0.408 | 0.476 | 0.48 | 0.48 | 0.639 |
| 1993 | 0.154 | 0.594 | 0.657 | 0.608 | 0.511 | 0.515 | 0.453 | 0.495 | 0.495 | 0.593 |
| 1994 | 0.162 | 0.548 | 0.804 | 0.537 | 0.481 | 0.397 | 0.32 | 0.401 | 0.401 | 0.593 |
| 1995 | 0.159 | 0.59 | 0.726 | 0.581 | 0.472 | 0.453 | 0.47 | 0.467 | 0.467 | 0.592 |
| 1996 | 0.181 | 0.517 | 0.636 | 0.668 | 0.528 | 0.414 | 0.577 | 0.509 | 0.509 | 0.587 |
| 1997 | 0.17 | 0.646 | 0.687 | 0.63 | 0.531 | 0.666 | 0.361 | 0.522 | 0.522 | 0.623 |
| 1998 | 0.065 | 0.446 | 0.709 | 0.508 | 0.442 | 0.392 | 0.367 | 0.402 | 0.402 | 0.526 |
| 1999 | 0.172 | 0.356 | 0.655 | 0.559 | 0.58 | 0.458 | 0.425 | 0.49 | 0.49 | 0.538 |
| 2000 | 0.155 | 0.473 | 0.615 | 0.617 | 0.472 | 0.457 | 0.286 | 0.406 | 0.406 | 0.544 |
| 2001 | 0.142 | 0.548 | 0.624 | 0.593 | 0.331 | 0.351 | 0.389 | 0.358 | 0.358 | 0.524 |
| 2002 | 0.325 | 0.576 | 0.659 | 0.671 | 0.577 | 0.39 | 0.293 | 0.421 | 0.421 | 0.621 |
| 2003 | 0.215 | 0.528 | 0.498 | 0.537 | 0.607 | 0.605 | 0.347 | 0.522 | 0.522 | 0.543 |
| 2004 | 0.175 | 0.538 | 0.619 | 0.622 | 0.641 | 0.54 | 0.566 | 0.538 | 0.538 | 0.605 |
| 2005 | 0.22 | 0.518 | 0.61 | 0.639 | 0.545 | 0.634 | 0.78 | 0.619 | 0.619 | 0.578 |
| 2006 | 0.31 | 0.642 | 0.597 | 0.623 | 0.545 | 0.415 | 0.76 | 0.661 | 0.661 | 0.602 |


| year/age | Fishing mortality-at-age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | $F(3-6)$ |
| 2007 | 0.172 | 0.586 | 0.738 | 0.717 | 0.716 | 0.713 | 0.351 | 0.652 | 0.652 | 0.689 |
| 2008 | 0.212 | 0.476 | 0.645 | 0.608 | 0.621 | 0.728 | 0.735 | 0.264 | 0.264 | 0.588 |
| 2009 | 0.154 | 0.451 | 0.343 | 0.315 | 0.34 | 0.326 | 0.334 | 0.321 | 0.321 | 0.362 |
| 2010 | 0.122 | 0.333 | 0.491 | 0.407 | 0.379 | 0.314 | 0.243 | 0.511 | 0.511 | 0.403 |
| 2011 | 0.124 | 0.325 | 0.403 | 0.275 | 0.22 | 0.259 | 0.273 | 0.255 | 0.255 | 0.306 |
| 2012 | 0.022 | 0.229 | 0.385 | 0.339 | 0.437 | 0.47 | 0.416 | 0.327 | 0.327 | 0.347 |
| 2013 | 0.022 | 0.143 | 0.295 | 0.292 | 0.28 | 0.454 | 0.357 | 0.806 | 0.806 | 0.253 |
| 2014 | 0.02 | 0.124 | 0.257 | 0.271 | 0.311 | 0.283 | 0.362 | 0.87 | 0.87 | 0.241 |
| 2015 | 0.009 | 0.07 | 0.162 | 0.245 | 0.261 | 0.315 | 0.182 | 0.433 | 0.433 | 0.185 |
| 2016 | 0.013 | 0.101 | 0.265 | 0.296 | 0.362 | 0.3 | 0.323 | 0.269 | 0.269 | 0.256 |
| 2017 | 0.024 | 0.115 | 0.247 | 0.433 | 0.362 | 0.296 | 0.228 | 0.461 | 0.461 | 0.289 |
| 2018 | 0.024 | 0.128 | 0.318 | 0.295 | 0.478 | 0.298 | 0.208 | 0.175 | 0.175 | 0.305 |

Table 23.7. Plaice in 7.e. Stock numbers-at-age. Numbers are rounded to the nearest thousand.

| year/age | Stock numbers-at-age [thousands] |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | sum |
| 1980 | 7067 | 2350 | 707 | 696 | 122 | 199 | 82 | 36 | 364 | 11623 |
| 1981 | 6961 | 5558 | 1371 | 397 | 404 | 50 | 118 | 52 | 265 | 15175 |
| 1982 | 3004 | 5545 | 2981 | 693 | 241 | 263 | 26 | 61 | 266 | 13080 |
| 1983 | 6382 | 2402 | 3102 | 1353 | 372 | 132 | 156 | 8 | 335 | 14243 |
| 1984 | 5788 | 4982 | 1378 | 1232 | 648 | 235 | 79 | 94 | 143 | 14579 |
| 1985 | 6959 | 4260 | 2865 | 601 | 605 | 395 | 123 | 44 | 154 | 16006 |
| 1986 | 7233 | 5611 | 2437 | 1293 | 388 | 303 | 218 | 84 | 180 | 17748 |
| 1987 | 14731 | 5555 | 2786 | 1306 | 697 | 240 | 150 | 131 | 151 | 25747 |
| 1988 | 12070 | 12064 | 2965 | 1194 | 548 | 403 | 132 | 87 | 186 | 29650 |
| 1989 | 8717 | 8994 | 6342 | 1606 | 630 | 346 | 253 | 82 | 259 | 27228 |
| 1990 | 3645 | 7478 | 5389 | 2920 | 647 | 306 | 195 | 136 | 172 | 20887 |
| 1991 | 3916 | 2922 | 3665 | 2267 | 1461 | 326 | 147 | 102 | 156 | 14961 |
| 1992 | 4420 | 2949 | 1469 | 1484 | 1110 | 816 | 192 | 89 | 137 | 12666 |
| 1993 | 4976 | 3262 | 1391 | 611 | 708 | 569 | 481 | 106 | 229 | 12334 |
| 1994 | 2311 | 3783 | 1597 | 639 | 295 | 377 | 302 | 271 | 200 | 9775 |
| 1995 | 2322 | 1743 | 1939 | 634 | 332 | 162 | 225 | 194 | 249 | 7799 |
| 1996 | 6722 | 1757 | 857 | 832 | 314 | 183 | 91 | 125 | 279 | 11160 |
| 1997 | 5846 | 4973 | 929 | 402 | 379 | 165 | 108 | 45 | 257 | 13104 |
| 1998 | 9109 | 4374 | 2313 | 415 | 190 | 197 | 75 | 66 | 194 | 16933 |
| 1999 | 4367 | 7574 | 2484 | 1009 | 221 | 108 | 118 | 46 | 122 | 16050 |
| 2000 | 2598 | 3261 | 4707 | 1144 | 512 | 110 | 61 | 69 | 165 | 12626 |
| 2001 | 3750 | 1974 | 1803 | 2256 | 548 | 283 | 62 | 40 | 131 | 10847 |
| 2002 | 3891 | 2885 | 1012 | 857 | 1106 | 349 | 177 | 37 | 124 | 10438 |
| 2003 | 4865 | 2493 | 1438 | 465 | 388 | 551 | 210 | 117 | 100 | 10627 |
| 2004 | 3105 | 3480 | 1303 | 775 | 241 | 188 | 267 | 131 | 97 | 9587 |
| 2005 | 4279 | 2311 | 1802 | 623 | 369 | 113 | 97 | 134 | 110 | 9837 |
| 2006 | 3965 | 3046 | 1220 | 868 | 291 | 190 | 53 | 39 | 116 | 9788 |


| year/age | Stock numbers-at-age [thousands] |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | sum |
| 2007 | 2641 | 2579 | 1422 | 596 | 413 | 150 | 111 | 22 | 110 | 8044 |
| 2008 | 5116 | 1971 | 1273 | 603 | 258 | 179 | 65 | 69 | 134 | 9669 |
| 2009 | 4804 | 3673 | 1086 | 592 | 291 | 123 | 77 | 28 | 65 | 10738 |
| 2010 | 7030 | 3652 | 2075 | 684 | 383 | 184 | 79 | 49 | 60 | 14196 |
| 2011 | 10343 | 5521 | 2322 | 1127 | 404 | 233 | 119 | 55 | 86 | 20209 |
| 2012 | 9879 | 8107 | 3540 | 1377 | 759 | 287 | 159 | 80 | 132 | 24321 |
| 2013 | 6689 | 8570 | 5721 | 2136 | 870 | 435 | 159 | 93 | 68 | 24741 |
| 2014 | 7089 | 5803 | 6589 | 3777 | 1414 | 583 | 245 | 99 | 64 | 25662 |
| 2015 | 11137 | 6160 | 4548 | 4518 | 2554 | 919 | 390 | 151 | 106 | 30483 |
| 2016 | 5136 | 9792 | 5094 | 3429 | 3137 | 1744 | 595 | 288 | 245 | 29461 |
| 2017 | 4007 | 4497 | 7849 | 3467 | 2261 | 1938 | 1146 | 382 | 384 | 25931 |
| 2018 | 3236 | 3471 | 3555 | 5436 | 1994 | 1397 | 1278 | 809 | 751 | 21927 |

Table 23.8. Plaice in 7.e. Assessment summary (raw values, not standardised).

| Year | Recruitment (age 2) [thousands] | TSB [tonnes] | SSB [tonnes] | Landings [TONNES] | Fbar(3-6) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 7067 | 4428 | 2512 | 1178 | 0.52 |
| 1981 | 6961 | 5034 | 2983 | 1676 | 0.44 |
| 1982 | 3004 | 4912 | 3252 | 1878 | 0.53 |
| 1983 | 6382 | 4410 | 3139 | 1714 | 0.55 |
| 1984 | 5788 | 4801 | 3135 | 1758 | 0.53 |
| 1985 | 6959 | 4759 | 3189 | 1677 | 0.50 |
| 1986 | 7233 | 5858 | 3747 | 2078 | 0.49 |
| 1987 | 14731 | 8193 | 4366 | 2272 | 0.60 |
| 1988 | 12070 | 8096 | 4746 | 2835 | 0.47 |
| 1989 | 8717 | 7805 | 5153 | 2742 | 0.61 |
| 1990 | 3645 | 6959 | 4987 | 2985 | 0.62 |
| 1991 | 3916 | 5283 | 4059 | 2183 | 0.60 |
| 1992 | 4420 | 4810 | 3447 | 1882 | 0.64 |
| 1993 | 4976 | 4360 | 2985 | 1614 | 0.59 |
| 1994 | 2311 | 3570 | 2555 | 1404 | 0.59 |
| 1995 | 2322 | 3104 | 2308 | 1247 | 0.59 |
| 1996 | 6722 | 3664 | 2266 | 1266 | 0.59 |
| 1997 | 5846 | 4195 | 2384 | 1583 | 0.62 |
| 1998 | 9109 | 3974 | 2463 | 1346 | 0.53 |
| 1999 | 4367 | 4270 | 2726 | 1543 | 0.54 |
| 2000 | 2598 | 3878 | 2957 | 1626 | 0.54 |
| 2001 | 3750 | 3481 | 2655 | 1310 | 0.52 |
| 2002 | 3891 | 3502 | 2437 | 1472 | 0.62 |
| 2003 | 4865 | 3710 | 2485 | 1387 | 0.54 |
| 2004 | 3105 | 3522 | 2276 | 1337 | 0.61 |
| 2005 | 4279 | 3477 | 2197 | 1319 | 0.58 |
| 2006 | 3965 | 3413 | 2091 | 1411 | 0.60 |
| 2007 | 2641 | 2668 | 1805 | 1146 | 0.69 |


| Year | Recruitment (age 2) [thousands] | TSB [tonnes] | SSB [tonnes] | Landings [TONNES] | Fbar(3-6) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 5116 | 3228 | 1928 | 1112 | 0.59 |
| 2009 | 4804 | 3290 | 2066 | 1024 | 0.36 |
| 2010 | 7030 | 4524 | 2656 | 1207 | 0.4 |
| 2011 | 10343 | 5785 | 3448 | 1417 | 0.31 |
| 2012 | 9879 | 7281 | 4414 | 1492 | 0.35 |
| 2013 | 6689 | 7824 | 5253 | 1472 | 0.25 |
| 2014 | 7089 | 7567 | 5693 | 1490 | 0.24 |
| 2015 | 11137 | 9982 | 7166 | 1424 | 0.185 |
| 2016 | 5136 | 10730 | 8081 | 2013 | 0.26 |
| 2017 | 4007 | 9476 | 7739 | 2128 | 0.29 |
| 2018 | 3236 | 8381 | 7352 | 1880 | 0.30 |
| 2019 | 5302* | 8367 | 6938 |  |  |

* geometric mean of time-series.

Table 23.9. Plaice in 7.e. Assessment summary (relative values).

| Year | Recruitment (age 2) <br> [relative] | SSB <br> [relative] | Landings <br> [TONNES] | DISCARDS [TONNES] | $F_{\text {bar }}(3-6)$ <br> [relative] |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 1.2 | 0.68 | 1178 |  | 1.04 |
| 1981 | 1.18 | 0.81 | 1676 |  | 0.88 |
| 1982 | 0.51 | 0.88 | 1878 |  | 1.06 |
| 1983 | 1.08 | 0.85 | 1714 |  | 1.11 |
| 1984 | 0.98 | 0.85 | 1758 |  | 1.06 |
| 1985 | 1.18 | 0.86 | 1677 |  | 1.01 |
| 1986 | 1.23 | 1.01 | 2078 |  | 0.98 |
| 1987 | 2.5 | 1.18 | 2272 |  | 1.21 |
| 1988 | 2.0 | 1.28 | 2835 |  | 0.94 |
| 1989 | 1.48 | 1.39 | 2742 |  | 1.23 |
| 1990 | 0.62 | 1.35 | 2985 |  | 1.25 |
| 1991 | 0.66 | 1.10 | 2183 |  | 1.21 |
| 1992 | 0.75 | 0.93 | 1882 |  | 1.29 |
| 1993 | 0.84 | 0.81 | 1614 |  | 1.19 |
| 1994 | 0.39 | 0.69 | 1404 |  | 1.19 |
| 1995 | 0.39 | 0.62 | 1247 |  | 1.19 |
| 1996 | 1.14 | 0.61 | 1266 |  | 1.18 |
| 1997 | 0.99 | 0.64 | 1583 |  | 1.25 |
| 1998 | 1.54 | 0.67 | 1346 |  | 1.06 |
| 1999 | 0.74 | 0.74 | 1543 |  | 1.08 |
| 2000 | 0.44 | 0.80 | 1626 |  | 1.10 |
| 2001 | 0.64 | 0.72 | 1310 |  | 1.05 |
| 2002 | 0.66 | 0.66 | 1472 |  | 1.25 |
| 2003 | 0.82 | 0.67 | 1387 |  | 1.09 |
| 2004 | 0.53 | 0.62 | 1337 |  | 1.22 |
| 2005 | 0.73 | 0.59 | 1319 |  | 1.16 |
| 2006 | 0.67 | 0.57 | 1411 |  | 1.21 |


| Year | Recruitment <br> (age 2) <br> [relative] | SSB <br> [relative] | Landings <br> [TONNES] | DISCARDS [TONNES] | $F_{\text {bar }}(3-6)$ <br> [relative] |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 0.45 | 0.49 | 1146 |  | 1.39 |
| 2008 | 0.87 | 0.52 | 1112 |  | 1.18 |
| 2009 | 0.81 | 0.56 | 1024 |  | 0.73 |
| 2010 | 1.19 | 0.72 | 1207 |  | 0.81 |
| 2011 | 1.75 | 0.93 | 1417 |  | 0.62 |
| 2012 | 1.67 | 1.19 | 1492 | 448 | 0.70 |
| 2013 | 1.13 | 1.42 | 1472 | 351 | 0.51 |
| 2014 | 1.20 | 1.54 | 1490 | 1133 | 0.48 |
| 2015 | 1.89 | 1.94 | 1424 | 1276 | 0.37 |
| 2016 | 0.87 | 2.2 | 2013 | 618 | 0.52 |
| 2017 | 0.68 | 2.1 | 2128 | 821 | 0.58 |
| 2018 | 0.55 | 1.99 | 1880 | 624 | 0.61 |
| 2019 | 0.90* | 1.87 |  |  |  |

* geometric mean of time-series (before standardisation).

Table 23.10. Plaice in 7e. The basis for the catch options for 2020. Note that one catch option is provided for stocks in ICES data categories $3-6$. The values presented here are the values presented during the working group.

| Division 7.e plaice stock |  |
| :---: | :---: |
| Index A (2018-2019) | 1.93 |
| Index B (2015-2017) | 2.1 |
| Index ratio (A/B) | 0.93 |
| Uncertainty cap | Not applied |
| Advised catch for 2019 issued 2018 | 3648 tonnes |
| Discard rate (2012-2018) | 0.30 |
| Precautionary buffer | Applied 0.8 |
| Catch advice** | 2721 tonnes |
| Wanted catch corresponding to catch advice^ | 1909 tonnes |
| \% Advice change (plaice Division 7.e stock)*** | -25 |
| Plaice in Division 7.e |  |
| Proportion of Division 7.e stock landings taken in Division 7.e (2003-2018) | 0.90 |
| Catch of plaice in Division 7.e corresponding to the advice for the stock | 2456 tonnes |
| Wanted catch of plaice in Division 7.e corresponding to the advice for the stock | 1722 tonnes |
| * The figures in the table are rounded. Calculations were done with unrounded inputs and computed values may not match exactly when calculated using the rounded figures in the table. |  |
| ** [recent advised catch] $\times$ [index ratio] x [precautionary buffer]. |  |
| *** Advice value 2020 relative to the advice value 2019. |  |
| $\wedge$ "Wanted catch" is used to describe fish that would be landed in the absence of the EU landing obligation. |  |

Table 23.11. Plaice in 7e. Results of the SPiCT model fit.

Convergence: O MSG: relative convergence (4)
ObJective function at optimum: 0.3165488
Euler time step (years) : $1 / 16$ or 0.0625
Nobs C: 39, Nobs I1: 13, Nobs I2: 16

Residual diagnostics (p-values)

|  | SHAPIRO | BIAS | ACF | LBOX SHAPIRO BIAS ACF | LBOX |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| C | 0.1222 | 0.6342 | 0.2916 | 0.4450 | - | - | - | - |
| I1 | 0.7154 | 0.7663 | 0.4470 | 0.7332 | - | - | - | - |
| I2 | 0.6115 | 0.6866 | 0.1542 | 0.3815 | - | - | - | - |

PRIORS

| LOGN | $\sim$ DNORM[LOG (2), 2 ^2] |
| ---: | :--- | :--- |
| LOGALPHA | $\sim$ DNORM[LOG(1), $\left.2^{\wedge} 2\right]$ |
| LOGBETA | $\sim \operatorname{DNORM[LOG(1),~2\wedge 2]~}$ |

Model parameter estimates w 95\% CI
ESTIMATE CILOW CIUPP LOG.EST
ALPHA1 4.350903E+00 0.9472487 1.998457E+01 1.4703833

ALPHA2 $2.563373 \mathrm{E}+00 \quad 0.48439591 .356511 \mathrm{E}+01 \quad 0.9413241$
BETA $1.696890 \mathrm{E}-01 \quad 0.03859507 .460640 \mathrm{E}-01-1.7737877$
R $2.458745 \mathrm{E}-01 \quad 0.05442561 .110770 \mathrm{E}+00-1.4029339$
RC $\quad 1.142545 \mathrm{E}+00 \quad 0.39090063 .339489 \mathrm{E}+00 \quad 0.1332578$
ROLD 4.316604E-01 0.0666692 2.794852E+00-0.8401162
M $\quad 2.145554 \mathrm{E}+031642.32726342 .802975 \mathrm{E}+03 \quad 7.6711531$
$\begin{array}{llllll}\mathrm{K} & 1.649977 \mathrm{E}+04 & 3391.5439837 & 8.027095 \mathrm{E}+04 & 9.7111018\end{array}$
Q1 $2.462392 \mathrm{E}-01 \quad 0.13474004 .500055 \mathrm{E}-01-1.4014520$
Q2 8.120000 $-04 \quad 0.00045601 .446000 \mathrm{E}-03-7.1159624$
N $\quad 4.303982 \mathrm{E}-01 \quad 0.25454297 .277459 \mathrm{E}-01-0.8430445$
SDB $\quad 8.067320 \mathrm{E}-02 \quad 0.01836043 .544668 \mathrm{E}-01$-2.5173493
SDF $2.049607 \mathrm{E}-01 \quad 0.14809902 .836540 \mathrm{E}-01-1.5849372$
SDI1 3.510011E-01 0.2296525 5.364704E-01 -1.0469660
SDI2 2.067954E-01 0.1281550 3.336924E-01 -1.5760252
SDC $\quad 3.477960 \mathrm{E}-02 \quad 0.00889491 .359902 \mathrm{E}-01-3.3587249$

Deterministic reference points (Drp)

|  | ESTIMATE | CILOW | CIUPP | LOG.EST |
| :--- | :---: | :---: | :---: | :---: |
| BMSYD | 3755.7468861 | 1009.8284196 | 13968.347888 | 8.2310424 |
| FMSYD | 0.5712723 | 0.1954503 | 1.669744 | -0.5598894 |
| MSYD | 2145.5539830 | 1642.3272634 | 2802.974777 | 7.6711531 |

Stochastic Reference points (SRp)

|  | ESTIMATE | CILOW | CIUPP | LOG.EST REL.DIFF.DRP |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| BMSYS | 3744.1834247 | 1009.8653832 | 13881.958675 | 8.2279588 | -0.003088380 |
| FMSYS | 0.5720508 | 0.1962089 | 1.667825 | -0.5585274 | 0.001360976 |
| MSYS | 2141.8723080 | 1641.4053273 | 2794.932432 | 7.6694356 | -0.001718905 |

States w 95\% CI (InP\$MSYtype: s)
ESTIMATE CILOW CIUPP LOG.EST

| B_2018.75 | 1181.2046995 | 586.9873935 | 2376.958275 | 7.0742901 |
| :--- | :---: | :---: | :---: | :---: | :---: |
| F_2018.75 | 1.5472271 | 0.7711580 | 3.104307 | 0.4364643 |

B_2018.75/BMSy $0.31547730 .0751756 \quad 1.323912-1.1536687$

| F_2018.75/FmSY | Y 2.7047020 | 0.8160355 | 8.964577 | 0.9949917 |
| :---: | :---: | :---: | :---: | :---: |
| Predictions w 95\% CI (inp\$msytype: s) |  |  |  |  |
|  | PREDICTION | CILOW C | CIUPP LOG.EST |  |
| B_2019.00 | 1169.720514 | 560.3095162 | 2441.946892 | 7.0645201 |
| F_2019.00 | 1.545162 | 0.7484009 | 3.190170 | 0.4351289 |
| B_2019.00/Bmsy | $Y \quad 0.312410$ | 0.0726658 | 1.343136 | -1.1634387 |
| F_2019.00/Fmsy | Y 2.701093 | 0.8003929 | 9.115399 | 0.9936563 |
| САтсн_2019.00 | 1788.193648 | 1459.6187495 | 52190.734069 | 7.4889613 |
| E (B_INF) 11 | 1141.814724 | NA | NA 7.0 | 0403741 |

### 24.14 Figures



Figure 23.1. Plaice in 7.e. International landings and discards by country as extracted from InterCatch for 20122018.


Figure 23.2. Plaice in 7.e. International landings and discards reported to InterCatch per country and fleet for the y ears 2012-2018.


Figure 23.3. Plaice in 7.e. Discard ratios for 2012-2018. "Fleet mean" is the mean of the ratios for all fleets which reported discards, "reported" is the proportion of reported discards in the reported catches, "weighted fleet mean" is the mean of the ratios for all fleets which reported discards weighted by the catch of the individual fleets, "raised" is the proportion of the discards as raised within InterCatch in the total catch for 7.e and "raised incl. migration" includes the catch (discards and landings) from Division 7.d used in the migration correction.


Figure 23.4. Plaice in 7.e. Landings, Discards and discard rate.


Figure 23.5. UK commercial lpue time-series. Lpue values are only shown for historical reasons but were not used in the assessment.


Figure 23.6. Plaice in 7.e. Length samples from InterCatch. The numbers are raised to fleet level.








Catch.Cat. $\square$ Discards Landings

Figure 23.6. Continued. Plaice in 7.e. Length samples from InterCatch. The numbers are raised to fleet level.


Figure 23.7. Plaice in 7.e. Total international length frequencies for 2014-2018 as raised within InterCatch for landings and discards including Length of first capture (Lc, calculated as first length class where the abundance is bigger or equal to half of maximum abundance) and mean length in the catch (Lmean, mean length above Lc).


Figure 23.8. Plaice in 7.e. Age-length key derived from samples from commercial UK fishery, split by landings and discards and including fit of von Bertalanffy growth function curves (red for 2017, blue for 2018 and black for all years combined).


Figure 23.9. Plaice in 7.e. Age samples from InterCatch. The numbers are raised to fleet level.


Figure 23.10. Plaice in 7.e. Landings age distribution.





Figure 23.11. Plaice in 7.e. Results of fitting a SPiCT model to the plaice 7.e stock.


Figure 23.12. Plaice in 7.e. Diagnostic plots of the SPiCT fit.


Figure 23.13. Plaice in 7.e. Retrospective and inverted retrospective SPiCT analysis.


Figure 23.14. Plaice in 7.e. Scientific tuning information used in the assessment including sum over all ages (right side).


Figure 23.15. Plaice in 7.e. Scientific tuning information used in the assessment standardised and cohort wise.


Figure 23.16. Plaice in 7.e. Internal consistency of the two survey time-series including correlation analysis.


Figure 23.17. Plaice in 7.e. Derivation of the 2018 stock and catch weights by applying a polynomial model to the raw InterCatch weights-at-age.


Figure 23.18. Plaice in 7.e. Landings and stock weights-at-age used in the assessment.


Figure 23.19. Plaice in 7.e. Landings and discards of the plaice 7.e stock disaggregated by the 7.e and the migration component from 7.d. Discard data are only available starting from 2012 for the Division 7.e.


Figure 23.20. Plaice in 7.e. XSA survey $\log$ catchability residuals.


Figure 23.21. Plaice in 7.e. Five-year retrospective of recruitment, spawning-stock biomass and fishing mortality estimates.


Figure 23.22. Plaice in 7.e. Summary of XSA final assessment. The plots on the top show the absolute values, the plots at the bottom the results relative to the mean of the time-series as used for the advice.


Figure 23.23. Plaice in 7.e. Comparison of the current XSA assessment run with the results from last year's WGCSE.


Figure 23.24. Plaice in 7.e. Total historical catches, split into landings and discards, including discard estimations prior to 2012.


Figure 23.25. Plaice in 7.e. Results of an exploratory total catch XSA assessment, in comparison to the landings only assessment and including reference points based on the total catch assessment.

# 25 Plaice (Pleuronectes platessa) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea) 

### 25.1 Type of assessment in 2019

Following the decision of the benchmark meeting WGFLAT 2011 the assessment of plaice stocks in ICES divisions 7.f-g should be carried out with analytic assessment model by Aarts and Poos (2009). This model derives relative trends, which include estimates of discards-at-age and was used for advice since 2012. In 2013 and 2015-2018 the AAP model had difficulties in interpreting the data due to conflicting trends between survey time-series and commercial time-series, particularly after 2010. The data have known issues in the recent years due to changes in effort reporting, as well as changes in discard practice. Therefore, the AAP was not used to provide advice at WGCSE 2015-2018 and advice was based instead on survey trends. As previous ICES advice used a catch/landings and biomass index series for the plaice in divisions 7.f-g, this stock dynamics was investigated in 2018-2019 by applying a biomass dynamic model (SPiCT-Stochastic Production model in Continuous Time), which provides model diagnostics. The diagnostics were found to be acceptable and therefore SPiCT was used as the basis for advice as a support for survey trends. ${ }^{1}$

## ICES advice applicable to 2019

Based on the ICES approach for data-limited stocks, ICES advises that catches in 2019 should be no more than 2160 tonnes.

## General stock description and management units

A TAC is allocated to ICES areas 7.f-g which corresponds to the stock area.
Management applicable to 2018 and 2019
TACs and quotas set for 2018 (source COUNCIL REGULATION (EU) No 120/2018)
Species: Plaice Pleuronectes platessa, Zone: 7.f and 7.g (PLE/7FG.)
Belgium 82
France 148
Ireland 204
United Kingdom 77
Total EU 511
Total TAC 511

[^15]TACs and quotas set for 2019 (source COUNCIL REGULATION (EU) No 124/2019).
Species: Plaice Pleuronectes platessa, Zone: 7.f and 7.g (PLE/7FG.)
Belgium 378
France 684
Ireland 243
United Kingdom 357
Total EU 1662
Total TAC 1662

TAC in 2019 are much higher than in 2018, as from now on they include fish that before was discarded.

## Fishery in 2018

As usual, the main fishery was concentrated on the Trevose Head ground off the north Cornish coast and around Land's End. Plaice was harvested throughout the year, with most of the catch landed from Q2 and Q3. The fleets harvesting plaice in the Celtic Sea primarily involved vessels from Belgium, France, Ireland and the UK. In 2018 Belgium reported $48.3 \%$ of the landings, France $30.2 \%$, Ireland $12.0 \%$ and the UK 9.5\%. The contribution of individual countries to total landings was similar to 2013-2017. The Working Group estimated that total international landings for 2018 were $421.7 \mathrm{t}, \sim 17.5 \%$ below the TAC of 511 t (Table28.1). Discards were a significant component of catch ( $\sim 55 \%$ in 2018), with the available time-series extending from 2004 to 2018. Discards haveexceeded landings since 2006. Most of the catch (52.7\%) were taken by beam trawlers, and $42.4 \%$; by bottom otter trawlers. Other gears accounted for $4.9 \%$. Effort and lpue of fishing fleets are presented in Tables 28.2-28.4.

### 25.2 Data

## Landings

National landings data and estimates of total landings and discards used by the WG are given in Table 28.1.

## Discards

Prior to 2010 indications were that discard rates, although variable, were substantial in some fleets/periods. At the ICES WKFLAT (2010) meeting, discard data from the countries participating in the fishery was raised and collated to the total international level for first time, a process that will be continued annually. The total estimates (Table 28.1) confirm the perception of the ongoing significant level of discarding. During the assessment of 2019, the discard information was available as annual summaries for Belgium, and as quarterly information for Ireland, France and the UK. Wherever this information was absent, discards were raised based on similarity of gear and quarter /annual type of data. WG estimates of the level of discards available from 2004 show a steady increase in time to levels higher than landings since 2006; in 2007, a substantial increase occurred in the discarding by all fleets. This is followed by a return to the previously lower levels until 2011 after which discards always exceeded landings. Data from 2018 national discard sampling programmes are summarised in Figures 28.1-28.3.

## Biological information

Quarterly or annual age compositions for 2018 were available for Belgium, Ireland, UK(E+W), and France métiers all together representing approximately $77.9 \%$ of the total landings (Figure 28.3).

International landings and discard numbers-at-age in years for which both are available (20042018) are compared in Figure 28.3;in recent years discards considerably exceeds landing in numbers at most ages. A strong recruitment cohort that appeared first in 2012 as 2-year old, in 2015 attained the age of 5 y.o. and began to predominate in landings, being still important in 2017 as 7 y.o fish. The next moderately strong generation (2 y.o. in 2015) appeared in 2015 and in 2017 represented important part of both landings and discards being the most abundant age group in 2018. Numbers- and weights-at-age for landings, discards and the stock used in the assessment are presented in Tables 28.5-28.9.

## Landings weight-at-age

Historically, landings weights-at-age were constructed by fitting a quadratic smoother through the aggregated catch weights for each year. WKFLAT (2011) decided not to continue with this approach following concerns raised by WGCSE that poor fits of the quadratic smoothing curve were resulting in the youngest ages being estimated to have heavier weights than adjacent older ages. WKFLAT (2011) rejected the use of the polynomial smoother for weights-at-age and suggested that raw landings weights are used in future. Raw data back to 1995 was obtained by WKFLAT (2011) and used to update the catch weights and stock weights files (Tables 28.6 and 28.9).

## Discard weight-at-age

Discard length and weight-at-age raw data were available for $\mathrm{UK}(\mathrm{E}+\mathrm{W})$, Belgium and Ireland. The national weight-at-age matrices were averaged to a total international estimate by weighting the individual weights-at-age for each year, by the catch numbers-at-age for each year and age (Tables 28.7 and 28.8).

## Stock weight-at-age

Where discard estimates were available from 2004 onwards, a revised set of stock weights-at-age were calculated. The stock weights were derived from the total international landings weights-at-age and the discard weights-at-age averaged by numbers-at-age from the respective datasets. Prior to 2004, a revised set of stock weights-at-age based on international landings data was produced. Thesenew values were based on collected weight data with a SOP correction (Table 28.9).

## Natural mortality and maturity

Estimates of natural mortality ( 0.12 for all years and all ages from tagging studies) were based on the value estimated for IrishSea plaice. The maturity ogive is based on UK(E\&W) 7.f-g survey data for March 1993 and March 1994 (Pawson and Harley, 1997). This maturity ogive was produced in 1997 and applied to all years in the assessment. Data were not used in the current assessment as AP model provided unsatisfactory residuals, so SPiCT was used instead.

| Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Maturity | 0 | 0.26 | 0.52 | 0.86 | 1.00 |

## Surveys

Indices of abundance from the UK(E\&W)-BTS-Q3 beam trawl survey in 7.f and the Irish IBTS survey (IGFS-WIBTS-Q4) in 7.g are presented in Table 28.10. Both surveys show consistent trends of the stock increases and decreases (Figure 28.4). The UK(E\&W)-BTS-Q3 started in 1995 and was always used for tuning the AP model. The Irish Celtic Explorer IBTS survey (IGFS-WIBTS-Q4) time-series started in 2003 and was not used in earlier years. The both survey time-series were used for the stock trends based advice in the years 2015, 2016 and 2017 and for SPiCT in 2018.

## Commercial landings per unit of effort

Commercial indices of abundance from the different fisheries provide contradictory trends (Figures 28.5 and 28.6). It occurred because of varying discarding practices from 2011 onwards, when fishermen began to discard substantial numbers of fish of commercial size. Therefore, these lpues, regardless their precision and objectiveness, could not be considered as proxies for adult fish abundance.

During this assessment, data on landings age structure were used up to the year 2010 (inclusive) because of a significant increase in the number of fish above MLS being discarded by fishermen thereafter. Up to the year 2012, the bulk of annual discards (all fleets combined) consisted of 2 or 2-3 y.o. fish, and in 2013-2018 mostly 3-5 y.o. fish (Figure 28.3). The level of discarding of adult fish differs between national fleets operating by the same fishing gear (e.g. beam trawls with mesh size of 70-99 mm (Figure 28.1).
Historically, the commercial lpue data illustrate a general pattern of steep decline since the high levels in the early 1990s, followed by a more gradual decline in the late 1990s. Since 2000, lpue has been relatively stable at a low level with small increases in some métiers, notably, in Belgian beam trawlers, the most important harvesters of the stock (Figure 28.6). Overall, the lpue rates remain at a relatively low level compared to historic catch rates.

## Other relevant data

There were no early closures of the fishery for plaice in 2018. The misreporting of landings of this stock is not considered to be a problem. Recent research on discard survival in the English Channel revealed that discard mortality of adult plaice captured by beam trawl varied with season, fish size and other factors like vessel type (Revill et al., 2013; Depestele et al., 2014; Uhlmann et al., 2016 a,b) Therefore significant amounts (4 to 93\%, mostly $<50 \%$ in Belgian beam trawlers and mean $48 \%$ in French beam trawlers) might survive discarding which has been confirmed by several (3-15) days of observations in captivity (Depestele et al., 2014; Uhlmann et al., 2016 a). The survival estimate for the UK otter trawl fishery in the Western Channel was 47-63\% and for the trammelnet fishery $71-72 \%$. The discard survival was alsoestimated as 19-20\% for the North Sea UK otter trawl fishery and $4-15 \%$ in the Western Channel UK beam trawl fishery (Catchpole et al., 2015). Smaller undersized plaice that represent the bulk of discards are likely to have relatively higher mortality as with other flatfish species (review: Hendrikson, Nies, 2007). As discard survival is unknown it might be not adequately be taken into consideration. There is no formal mixed fishery analysis for this area, but plaice in 7.fg is considered to be primarily a bycatch of the targeted sole fishery, so changes in effort in the directed sole fishery as well as multiannual management measures (EU, 2019) will impact fishing mortality on plaice.

### 25.3 Stock assessment

## Assessment model

WKFLAT (2011) agreed that the model that will be used as a temporary basis for the assessment and provision of advice for the Celtic Sea plaice is AP model (Aarts and Poos, 2009). This was selected on the basis that it was the only model available to WKFLAT which reconstructs the historic discarding rates (derived from the survey dataseries).

WKFLAT (2011) concluded that:

1. Due to the change in estimated fishing mortality when discards are included within the model fit, discards should be retained within the assessment model structure.
2. Given that the time-series of discard data, to which the models are fitted, is short and that, consequently, therearelikely to be changes in the management estimates as discard data are added in subsequent years, no definitive model structure can be recommended at this stage in the development process.
3. The most flexible of the models TVS_PTVS should be used as the basis for advice; in terms of relative changes in estimated total fishing mortality and biomass.
4. The other two models which provide similar structures should continue to be fitted at the WG to provide sensitivity comparisons.
5. As the dataseries are extended, a final model selection can be then determined.

In 2013, no assessment was presented for this stock given that the "preferred" Aarts and Poos (2009) model failed to converge and other model variants could not provide realistic representations of observed landings and discards. Consequently, WGCSE 2013 decided to avoid the use of the "preferred" TV_PTVS AP model variant and instead focus on assessing the stock using trends derived from the fishery-independent UK(E\&W) beam trawl survey. Trends derived from the UK (E\&W) beam trawl survey were selected for the basis of advice given that this survey most appropriately covered the spatial extent of the stock and well represented the mean age (25) landed in the fishery. The UK $(E \& W)$ beam trawl survey was used to infer trends in recruitment, stock size (spawning-stock biomass) and fishing mortality.

In 2014, corrected TV_PTVS Aarts and Poos (2009) model converged and produced realistic results and confirmed conclusions derived in 2013 from the fishery-independent UK(E\&W) beam trawl survey. In 2015-2017, all three model variants converged, but only of the "preferred" TV_PTVS AP variant provided estimations consistent with the previous run, observed catches and landings. However, trends of both UK(E\&W)-BTS-Q3 beam trawl and IGFS-WIBTS-Q4 surveys on one hand (Table 28.10) and data on lpues of commercial fleets (Table 28.11) produced conflicting signal that resulted in asymmetrical distribution of residuals. Because of this, the ICES stock advice was based on both surveys' cpue trends.

Independently of WGCSE, the stock status was explored in 2015 by WKLIFE using a biomass dynamic model (SPiCT) (ICES, 2016 a). As discard data were not available prior 2004, the group approximated the total catch values from 1977 to 2003. An adjustment was made to the data by applying the 2004 discard ratio back in time (landings prior to 2004 were multiplied by $\mathrm{K}=1.54$ ). These total catch data were combined with cpue trends of both surveys expressed in two meanstandardized biomass index series of +3 -year old plaice, which were considered to reflect "exploitable biomass" for this stock.
Results of modelling were found to be sensitive to truncating the catch to ensure $100 \%$ overlap between the survey and catch time-series. In this case, truncation lead to a $\sim 60 \%$ increase in $B_{\text {MSY }}$ and $\sim 30 \%$ decrease in FMSY, whereas CVs were hugely increased (by $\sim 200 \%$ and $\sim 75 \%$ respec-
tively). Therefore, the time-series was not truncated. Estimation of the observation error corresponding to the catch $(\beta)$ and survey $(\alpha)$ was tried, but the model did not converge when trying to estimate both of these, so $\alpha$ was fixed at 1 , while $\beta$ was estimated. Under all these assumptions the results indicated current stock status (2015) to be well above the biomass reference point 0.5 $B_{\text {MSY }}$, and F (2015) to be well be FMSY (ICES, 2016a).

In 2017, the ICES framework for category 3.2 stocks was applied (ICES, 2012;2016 b-d). As the previous ICES advice used both catch/landings and biomass index series, the stock was investigated by applying SPiCT. The SPiCT results were chosen to support the basis for advice using comparison of the two latest biomass index ( $\mathrm{B} / \mathrm{B}_{\mathrm{MSY}}$ ) values (index A ) with the three preceding values (index B), multiplied by the recent advised catch. The same approach was used in both 2018 and 2019.

## Final assessment

The settings and data for the model fits are set out in the table below the same way as in the previous year:

| ASSESSMENT YEAR |  | 2017 |
| :---: | :---: | :---: |
| Assessment model |  | SPiCT |
| Catch data |  | Including discards 1977-2017 (reported and raised discards for 20042017, and estimated discards for 1977-2003) |
| Discard rate |  | Average (proportion by number) 2004-2010. Calculated as discards/(landings + discards). |
| Tuning fleets | UK(E\&W)-BTS-Q3 | 1995-2017 ages 3+ |
|  | IGFS-WIBTS-Q4 | 2003-2017 ages 3+ |
|  | UK commercial beam trawl | 1993-2010 ages 4-8 |
|  | UK commercial otter trawl | 1993-2010 ages 4-8 |

Figure 28.7 presents the output plots for the model, and 28.8 it's diagnostics. Tables 28.12 and 28.13 contain information about the model diagnostics, deterministic and stochastic reference points and primary data of the model output.

## State of the stock

On the relative scale, the spawning biomass is estimated to have been increasing between 2005 and 2018 and began to decline in 2018, whereas $F$ has been steadily declining from 2001 onwards (Figure 28.11, Table 28.14). The estimated biomass was above Bmsy from 2013. Estimated F was below Fmsy from 2010, and upper limit of this estimation, from 2013. The observed stock increase was likely based on strong cohorts born in 2010 and 2013. The stock has been increasing from ~2008 after a period of low abundance in ~1995-2007. However, after three years of relatively low recruitment abundance in 2016-2018 (Figure 28.4) the stock size began to decline. As with other plaice stocks around the UK, like in the divisions $7 \mathrm{e}, 7 \mathrm{~h}-\mathrm{k}$ (ICES,2017) and North Sea (Dutz et al., 2016) this might be caused by some ongoing environmental changes.

### 25.4 Short-term projections

The short-term projection from the model for 2020 (Table 28.15) forecasts $\mathrm{B}>\mathrm{B}_{\text {MSY }}$ and $\mathrm{F}<\mathrm{F}_{\text {MSY }}$ within $95 \%$ confidence intervals.

### 25.5 Precautionary approach reference points

On the basis of the revision of the assessment datastructures and the AP model no MSY reference points were recommended for this stock. Meanwhile, using the SPiCT model at ICES WK Proxy (ICES, 2015) resulted in estimation of $B_{\text {trigger }}$ as $3800 t\left(50 \%\right.$ of $\left.B_{M S Y}\right)$ and $F_{M S Y}=0.27$. In 2019, application of the same model resulted in estimation of $B_{M S Y}$ as $4845 t$ and $F_{M S Y}=0.453$. A comparison of the two latest $B / B_{\text {msy }}$ index values with the three preceding values, multiplied by the recent advised catch demonstrated that estimated biomass to have increased by $\sim 6 \%$, so the uncertainty cap was not applied.

### 25.6 Management plans

The EU has proposed a multiannual management plan for the Western Waters (EU, 2018). However, this stock was excluded from the final version (EU 2019, approved on 05/03/2019 Meeting $n^{\circ} 3676$ - https://www.consilium.europa.eu/en/meetings/env/2019/03/05/). Therefore, there is no management plan for Celtic Sea plaice.

### 25.7 Uncertainties in assessment and forecast

## Landings

Sampling levels of landed catch (Figure 28.3) in recent years are sufficient to support current assessment approaches.

## Discards

Estimates of discarding are included in the assessment. From 2003 onwards, discard sampling for Ireland, Belgium, France and the UK(E\&W) has been improved under the Data Collection Regulation. Unknown levels of partial discard survival varying with fishing gear and season bring uncertainty into the assessment, which assumes that all discarded fish die. Discarding remains too high (exceeding landings) in this fishery, thereby compromising the effectiveness of quota management on landings. It is difficult to predict fishing fleet behaviour, as it is a commercial species of a low value taken mostly as a bycatch to fishery for sole, and to lesser extent, to Nephrops.

## Consistency

In 2015-2017, the advice for this stock was provided on the basis of research survey trends due to unreliability of the AP model results as well as conflicting trends between commercial vessels lpues (due to increasing discarding) and cpues of research surveys. In 2016-2018, the WGCSE decided to use results of the SPiCT model as a support source, output of which was consistent with trends in abundance of commercially sized fish aged 3+ as represented by data of research surveys.

### 25.8 References

Aarts, G., Poos, J.J. 2009. Comprehensive discard reconstruction and abundance estimation using flexible selectivity functions. ICES journal of marine science, 66: 763-771.
Catch pole, T., Randall, P., Forster, R., Smith, S., Ribeiro Santos, A., Armstrong, F., Hetherington, S., Bendall, V., Maxwell, D. 2015. Estimating the discard survival rates of selected commercial fish species (plaice Pleuronectes platessa) in four English fisheries (MF1234), Cefas report, pp 108.

Depestele J., Desender M., Benoît H.P., Polet H., Vincx M. 2014. Short-term survival of discarded target fish and non-target invertebrate species in the "eurocutter" beam trawl fishery of the southern North Sea. Fisheries Research 154: 82-92.

Dutz, J., Støttrup, Stenberg, C, Munk, P. 2016. Recent trends in the abundance of plaice Pleuronectes platessa and cod Gadus morhua in shallow coastal waters of the Northeastern Atlantic continental shelf - a review. Marine Biology Research 12: 785-796.

EU. 2018. Proposal for a REGULATION OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL establishing a multiannual plan for fish stocks in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amen ding Regulation (EU) 2016/1139 establishing a multiannual plan for the Baltic Sea, and repealing Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) 509/2007 and (EC) 1300/2008. COM/2018/0149 final. 30 pp. https://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:52018PC0149\&from=EN.

EU. 2019. REGULATION OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. 57 pp. https://data.consilium.europa.eu/doc/document/PE-78-2018-INIT/en/pdf.

Hendrikson, L., Nies, T. 2009. Discard and gear escapement survival rates of some Northeast groundfish species. NOAA Draft Working Paper. Data Meeting GARM 2008, October 29, 2007, 12 pp. (http://www.nefsc.noaa.gov/GARM-Public/1.DataMeeting/B.3\ Disc_survival_GARM2008.pdf).

Revill, A.S., Broadhurst, M.K., Millar, R.B. 2013. Mortality of adult plaice, Pleuronectes platessa and sole, Solea solea discarded from English Channel beam trawlers. Fisheries Research, 147: 320-326.
ICES. 2012. ICES Implementation of Advice for Data-limited Stocks in 2012 in its 2012 Advice. ICES CM 2012/ACOM 68.42 pp .
ICES. 2015. Report of the Workshop to consider MSY proxies for stocks in ICES category 3 and 4 stocks in Western Waters (WKProxy), 3-6 November 2015. ICES CM 2015/ACOM:61.

ICES. 2016a. Report of the Workshop to consider MSY proxies for stocks in ICES category 3 and 4 stocks in Western Waters (WKProxy), 3-6 November 2015, ICES Headquarters, Copenhagen. ICES CM 2015/ACOM:61. 183 pp.
ICES. 2016b. EU request to provide a framework for the classification of stock status relative to MSY proxies for selected category 3 and category 4 stocks in ICES subareas 5 to 10. In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.4.2.

ICES. 2016c. Advice basis. In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 1, Section 1.2.

ICES. 2016d. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 4-13 May 2016, ICES Headquarters, Copenhagen, Denmark. ICES CM 2016/ACOM:13.
ICES. 2017. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 9-18 May 2017, ICES Headquarters, Copenhagen, Denmark. ICES CM 2017/ACOM:13.

Uhlmann, S. S., Theunynck, R., Ampe, B., Verkempynck, R., Miller, D. C. M., van Marlen, B., van der Reijden, K., Molenaar, P., Vanderperren, E., Polet, H. 2016 a. Overleving door boomkor gevangen pladijs - survival of beam-trawled European plaice (Pleuronectes platessa). ILVO Mededeling 210. Institute for Agricultural and Fisheries Research, Oostende, Belgium. 172 pp.

Uhlmann, S. S., Theunynck, R., Ampe, B., Desender, M., Soetaert, M., and Depestele, J. Injury, reflex impairment, and survival of beam-trawled flatfish. 2016 b. ICES Journal of Marine Science, 73: 1244-1254.

Table 28.1. Plaice in divisions 7.f-g. Nominal landings ( $\mathbf{t}$ ) as reported to ICES, and total landings as used by ICES WG CSE

|  | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 214 | 196 | 171 | 372 | 365 | 341 | 314 | 283 | 357 | 665 |
| UK (Engl. \& Wales) | 150 | 152 | 176 | 227 | 251 | 196 | 279 | 366 | 466 | 529 |
| France | 365 | 527 | 467 | 706 | 697 | 568 | 532 | 558 | 493 | 878 |
| Ireland | 28 | 0 | 49 | 61 | 64 | 198 | 48 | 72 | 91 | 302 |
| N. Ireland |  |  |  |  |  |  |  |  |  |  |

Table 28.1. Plaice in divisions 7.f-g. Nominal landings ( $t$ ) as reported to ICES, and total landings as used by ICES WG CSE (continued).

|  | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 194 | 188 | 216 | 188 | 210 | 203 | 185 | 182 | 185 | 243 |
| UK | 61 | 63 | 55 | 54 | 45 | 45 | 41 | 25 | 25 | 27 |
| France | 104 | 70 | NA | 136 | 98 | 125 | 106 | 155 | 111 | 108 |
| Ireland | 58 | 64 | 63 | 63 | 67 | 76 | 80 | 49 | 59 | 52 |
| Total reported | 417 | 385 | NA | 442 | 420 | 450 | 412 | 411 | 381 | 430 |
| Discards | 1288 | 583 | 608 | 670 | 1107 | 1123 | 1274 | 772 | 778 | 571 |
| Unallocated | -7 | 52 | -1 | -9 | 7 | -8 | -2 | -1 | 0 | 0 |
| Landings used by WG | 410 | 437 | 481 | 442 | 427 | 442 | 414 | 410 | 381 | 431 |
| Catch as used by WG | 1698 | 1020 | 1089 | 1112 | 1534 | 1565 | 1688 | 1183 | 1159 | 1002 |
|  | 2017 | 2018 |  |  |  |  |  |  |  |  |
| Belgium | 179 | 204 |  |  |  |  |  |  |  |  |
| UK | 38 | 40 |  |  |  |  |  |  |  |  |
| France | 108 | 127 |  |  |  |  |  |  |  |  |
| Ireland | 63 | 51 |  |  |  |  |  |  |  |  |
| Total reported | 388 | 422 |  |  |  |  |  |  |  |  |
| Unallocated | 1284 | 930 |  |  |  |  |  |  |  |  |

Table 28.2. Plaice in divisions 7.f-g: lpue and cpue for UK(E\&W) fleets.

|  | LANDINGS PER UNIT EFFORT (LPUE) kglday |  |  |  |  |  |  |  |  |  | TRAWL |  | BEAM TRAWL |  | VIIfg EFFORT |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | RECT. <br> GROUP |  | RECT. <br> GROUP |  | VIlg EAST (grp 2) |  | RECT. <br> GROUP |  | VIIg WEST (grp 3) |  | RECT. GROUP |  | RECT. GROUP |  |  |  |
|  | VIlf (grp |  | VIIg EA <br> 2) | T (grp | Effort |  | VIlg WE <br> 3) | ST (grp | Effort |  | VIIf (grp 1) |  | VIlf (grp 1) |  | TRAWL | BEAM |
|  | Trawl | BEAM | Trawl | BEAM | TRAWL | BEAM | Trawl | BEAM | TRAWL | BEAM | LANDINGS | EFFORT | LANDINGS | EFFORT | (000) | (000) |
| Year |  | Trawl |  | Trawl | (Days fished) | (Days fished) |  | Trawl | (Days fished) | (Days fished) | (t) | (Days <br> fished) | (t) | (Days <br> fished) | (Days <br> fished) | (Days <br> fished) |
| 1983 | 86.39 | 30.33 | 71.84 | 54.85 | 82 | 149 | 0.00 | 75.69 | 0 | 8 | 53.96 | 620 | 5.62 | 195 | 702 | 353 |
| 1984 | 79.67 | 99.69 | 94.50 | $\begin{gathered} 106.6 \\ 5 \end{gathered}$ | 316 | 298 | 0.00 | 66.96 | 0 | 129 | 156.33 | 1723 | 99.01 | 901 | 2039 | 1328 |
| 1985 | $\begin{gathered} 115.9 \\ 3 \end{gathered}$ | $\begin{gathered} 122.9 \\ 1 \end{gathered}$ | $\begin{gathered} 119.6 \\ 3 \end{gathered}$ | $\begin{gathered} 174.3 \\ 9 \end{gathered}$ | 206 | 285 | 67.62 | 233.25 | 23 | 92 | 188.60 | 1493 | 146.71 | 1101 | 1722 | 1478 |
| 1986 | $\begin{gathered} 119.8 \\ 1 \end{gathered}$ | $\begin{gathered} 113.6 \\ 2 \end{gathered}$ | $\begin{gathered} 103.3 \\ 7 \end{gathered}$ | $\begin{gathered} 183.7 \\ 2 \end{gathered}$ | 334 | 180 | 49.93 | 380.20 | 35 | 29 | 138.48 | 1125 | 91.16 | 973 | 1494 | 1182 |
| 1987 | $\begin{gathered} 131.2 \\ 7 \end{gathered}$ | $\begin{gathered} 114.3 \\ 4 \end{gathered}$ | $\begin{gathered} 223.1 \\ 3 \end{gathered}$ | $\begin{gathered} 291.3 \\ 0 \end{gathered}$ | 364 | 187 | 33.68 | 446.46 | 26 | 26 | 196.01 | 1211 | 148.39 | 1681 | 1601 | 1894 |
| 1988 | $\begin{gathered} 232.5 \\ 1 \end{gathered}$ | $\begin{gathered} 247.9 \\ 1 \end{gathered}$ | $\begin{gathered} 217.1 \\ 1 \end{gathered}$ | $\begin{gathered} 356.0 \\ 2 \end{gathered}$ | 351 | 77 | 48.43 | 670.38 | 20 | 36 | 200.68 | 838 | 205.01 | 1102 | 1210 | 1215 |
| 1989 | $\begin{gathered} 130.8 \\ 4 \end{gathered}$ | $\begin{gathered} 138.6 \\ 2 \end{gathered}$ | $\begin{gathered} 137.7 \\ 6 \end{gathered}$ | $\begin{gathered} 293.8 \\ 9 \end{gathered}$ | 327 | 125 | 86.54 | 575.30 | 15 | 7 | 129.65 | 966 | 96.15 | 861 | 1309 | 994 |
| 1990 | 75.55 | 88.83 | 59.00 | $\begin{gathered} 166.6 \\ 9 \end{gathered}$ | 435 | 165 | 78.13 | 147.13 | 24 | 194 | 97.39 | 1229 | 155.84 | 1256 | 1689 | 1615 |


|  | LANDINGS PER UNIT EFFORT (LPUE) kglday |  |  |  |  |  |  |  |  |  | TRAWL |  | BEAM TRAWL |  | VIIfg EFFORT |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | RECT. GROUP |  | RECT. GROUP |  | VIlg EAS |  | RECT. GROUP |  | VIlg WE |  | RECT. G |  | RECT. GR |  |  |  |
| 1991 | 48.20 | 93.83 | 44.90 | 73.40 | 306 | 483 | 42.22 | 109.40 | 45 | 104 | 55.72 | 1066 | 190.79 | 1667 | 1417 | 2254 |
| 1992 | 49.33 | 57.20 | 41.29 | 69.80 | 303 | 633 | 45.00 | 70.04 | 435 | 90 | 44.92 | 898 | 91.34 | 1420 | 1636 | 2143 |
| 1993 | 43.85 | 69.98 | 23.83 | 65.14 | 251 | 694 | 56.64 | 32.85 | 30 | 135 | 38.41 | 836 | 109.37 | 1669 | 1117 | 2497 |
| 1994 | 39.67 | 40.41 | 31.76 | 49.39 | 225 | 610 | 10.70 | 70.61 | 19 | 116 | 23.21 | 623 | 86.14 | 2219 | 866 | 2945 |
| 1995 | 41.81 | 43.01 | 30.91 | 54.05 | 196 | 694 | 61.67 | 37.12 | 30 | 128 | 26.39 | 580 | 96.10 | 2303 | 807 | 3125 |
| 1996 | 38.80 | 33.67 | 26.25 | 27.49 | 341 | 560 | 6.15 | 11.82 | 105 | 220 | 23.68 | 593 | 81.19 | 2391 | 1038 | 3170 |
| 1997 | 34.61 | 31.01 | 21.37 | 33.42 | 370 | 770 | 17.47 | 7.50 | 122 | 146 | 20.76 | 577 | 85.13 | 2661 | 1069 | 3578 |
| 1998 | 21.86 | 26.07 | 15.53 | 15.33 | 385 | 591 | 5.12 | 12.65 | 94 | 159 | 10.97 | 517 | 85.15 | 2846 | 995 | 3597 |
| 1999 | 35.60 | 26.62 | 20.65 | 12.00 | 176 | 1461 | 5.14 | 11.96 | 235 | 312 | 12.06 | 395 | 85.55 | 3058 | 806 | 4831 |
| 2000 | 32.09 | 16.10 | 40.58 | 11.64 | 187 | 1007 | 3.35 | 10.10 | 160 | 200 | 10.99 | 284 | 53.59 | 3133 | 630 | 4341 |
| 2001 | 34.02 | 16.69 | 32.30 | 15.26 | 187 | 1155 | 4.66 | 11.04 | 179 | 91 | 9.82 | 309 | 53.47 | 3172 | 675 | 4418 |
| 2002 | 19.78 | 15.64 | 48.80 | 20.81 | 123 | 463 | 7.43 | 4.81 | 170 | 60 | 6.91 | 416 | 38.85 | 2652 | 709 | 3174 |
| 2003 | 23.45 | 18.24 | 8.19 | 20.78 | 51 | 772 | 4.48 | 1.49 | 124 | 158 | 15.85 | 696 | 50.94 | 2669 | 871 | 3599 |
| 2004 | 18.77 | 15.54 | 8.66 | 7.81 | 198 | 923 | 3.09 | 3.39 | 125 | 178 | 12.45 | 641 | 40.72 | 2503 | 965 | 3604 |
| 2005 | 11.20 | 11.00 | 2.14 | 8.25 | 21 | 618 | 0.25 | 1.33 | 154 | 116 | 9.55 | 876 | 23.25 | 1968 | 1051 | 2702 |
| 2006 | 21.21 | 12.77 | 5.91 | 15.19 | 23 | 630 | 0.64 | 0.58 | 233 | 70 | 19.94 | 924 | 14.31 | 1330 | 1181 | 2030 |
| 2007 | 14.79 | 17.93 | 20.42 | 10.58 | 31 | 518 | 1.71 | 5.90 | 219 | 12 | 12.09 | 798 | 18.18 | 1407 | 1048 | 1937 |



Table 28.3. Plaice in divisions 7.f-g: lpue and effort for Belgian fleets in 7.f-g.

| BELGIAN Beam Trawl 7fg |  |  |  |
| :---: | :---: | :---: | :---: |
| Year | Landings (t) | Effort (000 hr) | Ipue (kg/h) |
| 1996 | 356.89 | 53.27 | 6.70 |
| 1997 | 474.71 | 57.36 | 8.28 |
| 1998 | 443.38 | 57.79 | 7.67 |
| 1999 | 410.22 | 55.11 | 7.44 |
| 2000 | 230.63 | 51.34 | 4.49 |
| 2001 | 274.84 | 54.90 | 5.01 |
| 2002 | 259.80 | 49.60 | 5.24 |
| 2003 | 215.95 | 62.73 | 3.44 |
| 2004 | 207.27 | 78.73 | 2.63 |
| 2005 | 153.73 | 64.50 | 2.38 |
| 2006 | 134.44 | 50.28 | 2.67 |
| 2007 | 139.39 | 45.72 | 3.05 |
| 2008 | 106.29 | 28.71 | 3.70 |
| 2009 | 140.76 | 30.84 | 4.56 |
| 2010 | 127.15 | 32.74 | 3.88 |
| 2011 | 159.03 | 41.41 | 3.84 |
| 2012 | 165.73 | 46.25 | 3.58 |
| 2013 | 155.973 | 45.159 | 3.454 |
| 2014 | 155.317 | 31.271 | 4.967 |
| 2015 | 165.17 | 31.792 | 5.195 |
| 2016 | 212.01 | 32.34 | 6.556 |
| 2017 | 169.03 | 33.35 | 5.07 |
| 2018 | 186.861 | 31.48 | 5.94 |

Table 28.4. Plaice in Divisions 7.f-g: lpue and effort for Irish otter trawl, beam and seine fleets in 7.g.

| IR-OTB-7G |  |  |  | IR-SCC-7G |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | Landings (t) | Effort (000 hr) | lpue (kg/h) | Landings (t) | Effort (000 hr) | Ipue (kg/h) |
| 1995 | 94.23 | 63.56 | 1.48 | 9.55 | 6.43 | 1.49 |
| 1996 | 133.66 | 60.04 | 2.23 | 14.20 | 9.73 | 1.46 |
| 1997 | 119.84 | 65.10 | 1.84 | 38.79 | 16.13 | 2.40 |
| 1998 | 96.72 | 72.30 | 1.34 | 21.38 | 14.94 | 1.43 |
| 1999 | 60.05 | 51.66 | 1.16 | 10.40 | 8.01 | 1.30 |
| 2000 | 28.78 | 60.60 | 0.47 | 11.40 | 9.90 | 1.15 |
| 2001 | 23.82 | 69.43 | 0.34 | 10.93 | 16.33 | 0.67 |
| 2002 | 42.30 | 77.69 | 0.54 | 16.42 | 20.86 | 0.79 |
| 2003 | 26.35 | 86.79 | 0.30 | 13.80 | 20.91 | 0.66 |
| 2004 | 26.62 | 96.99 | 0.27 | 5.04 | 19.38 | 0.26 |
| 2005 | 22.78 | 124.40 | 0.18 | 6.47 | 14.81 | 0.44 |
| 2006 | 25.17 | 119.23 | 0.21 | 5.10 | 14.79 | 0.34 |
| 2007 | 30.99 | 136.52 | 0.23 | 4.76 | 15.82 | 0.30 |
| 2008 | 39.17 | 125.81 | 0.31 | 8.38 | 11.65 | 0.72 |
| 2009 | 43.81 | 137.11 | 0.32 | 7.98 | 8.19 | 0.98 |
| 2010 | 44.29 | 140.65 | 0.31 | 10.71 | 9.69 | 1.11 |
| 2011 | 44.68 | 120.33 | 0.37 | 11.12 | 11.01 | 1.01 |
| 2012 | 43.21 | 121.08 | 0.35 | 18.41 | 14.15 | 1.30 |
| 2013 | 31.91 | 118.13 | 0.28 | 11.10 | 12.06 | 0.84 |
| 2014 | 28.00 | 127.40 | 0.22 | 7.60 | 12.00 | 0.61 |
| 2015 | 33.34 | 132.69 | 0.25 | 8.36 | 9.28 | 0.90 |
| 2016 | 34.80 | 148.17 | 0.23 | 9.37 | 10.44 | 0.90 |
| 2017 | 40.86 | 135.98 | 0.30 | 10.49 | 9.75 | 1.08 |
| 2018 | 33.89 | 108.22 | 0.31 | 8.13 | 9.69 | 0.84 |
|  |  |  |  |  |  |  |

IR-TBB-7G

| Year | Landings (t) | Effort (000 hr) | Ipue (kg/h) | Year | Landings (t) | Effort (000 hr. Ipue (kg/h) |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1995 | 37.92 | 20.78 | 1.83 | 2007 | 21.18 | 55.86 | 0.38 |
| 1996 | 53.02 | 26.76 | 1.98 | 2008 | 14.18 | 37.22 | 0.38 |
| 1997 | 94.59 | 28.25 | 3.35 | 2009 | 6.96 | 37.96 | 0.18 |
| 1998 | 122.13 | 35.25 | 3.46 | 2010 | 6.56 | 40.22 | 0.16 |
| 1999 | 25.80 | 40.87 | 0.63 | 2011 | 6.71 | 35.33 | 0.19 |
| 2000 | 12.62 | 37.03 | 0.34 | 2012 | 33.63 | 40.33 | 0.83 |
| 2001 | 4.80 | 39.71 | 0.12 | 2013 | 32.32 | 38.48 | 0.84 |
| 2002 | 7.08 | 31.62 | 0.22 | 2014 | 12.50 | 37.80 | 0.33 |
| 2003 | 9.37 | 49.26 | 0.19 | 2015 | 12.10 | 37.79 | 0.32 |
| 2004 | 6.17 | 54.86 | 0.11 | 2016 | 9.83 | 39.55 | 0.25 |
| 2005 | 9.49 | 49.65 | 0.19 | 2017 | 12.39 | 35.21 | 0.35 |
| 2006 | 14.46 | 60.48 | 0.24 | 2018 | 9.62 | 37.42 | 0.26 |

Table 28.5. Plaice in divisions 7.f-g. Landings numbers-at-age.

| Landings numbers-at-age |  |  | Numbers*10**-3 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE\YEAR | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 989 | 851 | 877 | 1921 | 822 | 300 | 750 | 704 | 1461 | 703 |
| 3 | 426 | 903 | 673 | 1207 | 2111 | 1180 | 560 | 918 | 2503 | 2595 |
| 4 | 411 | 291 | 638 | 658 | 681 | 955 | 827 | 343 | 393 | 1332 |
| 5 | 105 | 136 | 72 | 146 | 109 | 443 | 372 | 373 | 102 | 156 |
| 6 | 72 | 76 | 70 | 21 | 54 | 86 | 92 | 209 | 177 | 59 |
| 7 | 37 | 47 | 34 | 16 | 53 | 51 | 44 | 70 | 62 | 48 |
| 8 | 59 | 23 | 8 | 16 | 11 | 14 | 27 | 41 | 25 | 32 |
| +gp | 75 | 98 | 46 | 32 | 44 | 60 | 23 | 42 | 38 | 24 |
| TOTALNUM | 2175 | 2426 | 2419 | 4018 | 3886 | 3090 | 2696 | 2701 | 4762 | 4950 |
| AGE\YEAR | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 25 | 100 | 43 | 0 |
| 2 | 434 | 967 | 797 | 164 | 279 | 800 | 1019 | 428 | 488 | 812 |
| 3 | 1883 | 2099 | 3550 | 2078 | 1072 | 526 | 1179 | 936 | 572 | 734 |
| 4 | 1812 | 1568 | 1807 | 2427 | 1193 | 357 | 284 | 730 | 743 | 515 |
| 5 | 772 | 612 | 741 | 655 | 578 | 471 | 139 | 164 | 334 | 219 |
| 6 | 156 | 413 | 160 | 242 | 179 | 275 | 185 | 117 | 117 | 137 |
| 7 | 22 | 65 | 98 | 86 | 94 | 80 | 115 | 86 | 57 | 59 |
| 8 | 125 | 16 | 24 | 70 | 78 | 21 | 62 | 92 | 48 | 37 |
| +gp | 76 | 73 | 23 | 46 | 79 | 96 | 59 | 65 | 132 | 96 |
| TOTALNUM | 5281 | 5814 | 7201 | 5769 | 3553 | 2627 | 3066 | 2716 | 2534 | 2609 |
| AGE\YEAR | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 8 | 17 | 22 | 19 | 75 | 3 | 15 | 6 | 24 | 12 |
| 2 | 420 | 426 | 243 | 320 | 651 | 170 | 239 | 126 | 201 | 331 |
| 3 | 1318 | 921 | 982 | 606 | 371 | 661 | 571 | 578 | 327 | 458 |
| 4 | 929 | 849 | 802 | 482 | 323 | 543 | 465 | 428 | 265 | 140 |
| 5 | 272 | 287 | 372 | 203 | 199 | 183 | 150 | 261 | 134 | 134 |


| Landings numbers-at-age |  | Numbers*10**-3 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 121 | 96 | 116 | 145 | 108 | 113 | 85 | 46 | 73 | 76 |
| 7 | 60 | 82 | 45 | 53 | 62 | 65 | 34 | 27 | 24 | 50 |
| 8 | 20 | 39 | 27 | 22 | 23 | 24 | 26 | 15 | 14 | 12 |
| +gp | 82 | 56 | 69 | 32 | 28 | 28 | 24 | 17 | 16 | 15 |
| TOTALNUM | 3231 | 2773 | 2678 | 1881 | 1838 | 1789 | 1608 | 1504 | 1078 | 1229 |
| AGE\YEAR | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| 1 | 8 | 15 | 2 | 3 | 1 | 2 | 3 | 0 | 0 | 2 |
| 2 | 130 | 270 | 127 | 135 | 135 | 106 | 64 | 24 | 55 | 20 |
| 3 | 513 | 341 | 626 | 223 | 326 | 485 | 328 | 123 | 122 | 332 |
| 4 | 340 | 443 | 345 | 430 | 208 | 288 | 383 | 452 | 231 | 201 |
| 5 | 104 | 145 | 273 | 191 | 248 | 164 | 192 | 247 | 410 | 182 |
| 6 | 76 | 47 | 68 | 152 | 130 | 163 | 67 | 109 | 127 | 228 |
| 7 | 46 | 29 | 20 | 44 | 69 | 65 | 70 | 33 | 43 | 94 |
| 8 | 26 | 11 | 10 | 8 | 28 | 33 | 29 | 36 | 17 | 42 |
| +gp | 13 | 15 | 12 | 8 | 17 | 23 | 31 | 30 | 26 | 37 |
| TOTALNUM | 1257 | 1315 | 1485 | 1187 | 1161 | 1329 | 1167 | 1054 | 1052 | 1138 |
| AGE\YEAR | 2017 | 2018 |  |  |  |  |  |  |  |  |
| 1 | 0 | 3 |  |  |  |  |  |  |  |  |
| 2 | 33 | 32 |  |  |  |  |  |  |  |  |
| 3 | 57 | 143 |  |  |  |  |  |  |  |  |
| 4 | 380 | 122 |  |  |  |  |  |  |  |  |
| 5 | 167 | 393 |  |  |  |  |  |  |  |  |
| 6 | 112 | 160 |  |  |  |  |  |  |  |  |
| 7 | 145 | 92 |  |  |  |  |  |  |  |  |
| 8 | 56 | 89 |  |  |  |  |  |  |  |  |
| +gp | 35 | 62 |  |  |  |  |  |  |  |  |
| TOTALNUM | 985 | 1096 |  |  |  |  |  |  |  |  |

Table 28.6. Plaice in divisions 7.f-g. Landings weights-at-age.

| Landings weights-at-age (kg) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE\YEAR | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 0.078 | 0.194 | 0.076 | 0.118 | 0.185 | 0.151 | 0.178 | 0.276 | 0.135 | 0.000 |
| 2 | 0.205 | 0.258 | 0.203 | 0.238 | 0.255 | 0.245 | 0.274 | 0.324 | 0.251 | 0.160 |
| 3 | 0.323 | 0.323 | 0.325 | 0.354 | 0.330 | 0.339 | 0.369 | 0.384 | 0.363 | 0.301 |
| 4 | 0.430 | 0.389 | 0.440 | 0.467 | 0.412 | 0.433 | 0.464 | 0.455 | 0.470 | 0.434 |
| 5 | 0.528 | 0.457 | 0.550 | 0.576 | 0.500 | 0.526 | 0.559 | 0.538 | 0.572 | 0.559 |
| 6 | 0.615 | 0.525 | 0.652 | 0.682 | 0.595 | 0.620 | 0.654 | 0.633 | 0.670 | 0.677 |
| 7 | 0.693 | 0.595 | 0.749 | 0.784 | 0.695 | 0.714 | 0.749 | 0.739 | 0.763 | 0.787 |
| 8 | 0.760 | 0.666 | 0.839 | 0.882 | 0.802 | 0.808 | 0.844 | 0.857 | 0.851 | 0.889 |
| +gp | 0.8762 | 0.8435 | 1.0653 | 1.1812 | 1.1824 | 1.0948 | 1.1579 | 1.2661 | 1.0036 | 1.1033 |
| SOPCOFAC | 1.0052 | 1.0262 | 1.0225 | 1.0135 | 1.0042 | 1.0125 | 0.9995 | 1.0000 | 1.0047 | 0.9997 |
| AGE\YEAR | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | 0.129 | 0.260 | 0.102 | 0.240 | 0.200 | 0.148 | 0.171 | 0.236 | 0.219 | 0.000 |
| 2 | 0.208 | 0.288 | 0.176 | 0.270 | 0.260 | 0.257 | 0.263 | 0.296 | 0.254 | 0.247 |
| 3 | 0.288 | 0.325 | 0.255 | 0.309 | 0.327 | 0.362 | 0.314 | 0.308 | 0.304 | 0.295 |
| 4 | 0.368 | 0.370 | 0.337 | 0.358 | 0.400 | 0.464 | 0.405 | 0.397 | 0.364 | 0.349 |
| 5 | 0.449 | 0.423 | 0.423 | 0.416 | 0.481 | 0.563 | 0.500 | 0.455 | 0.485 | 0.512 |
| 6 | 0.530 | 0.484 | 0.514 | 0.483 | 0.567 | 0.658 | 0.598 | 0.598 | 0.603 | 0.553 |
| 7 | 0.612 | 0.554 | 0.608 | 0.560 | 0.661 | 0.750 | 0.643 | 0.801 | 0.714 | 0.523 |
| 8 | 0.694 | 0.633 | 0.706 | 0.646 | 0.761 | 0.839 | 0.728 | 0.728 | 0.752 | 0.947 |
| +gp | 0.8632 | 0.8887 | 0.9932 | 0.9097 | 1.0465 | 1.0399 | 0.9886 | 0.9585 | 1.0655 | 1.0667 |
| SOPCOFAC | 1.0034 | 1.0024 | 1.0006 | 1.0009 | 1.0113 | 1.0022 | 0.9997 | 1.0001 | 1.0004 | 0.9998 |
| AGE\YEAR | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 0.249 | 0.213 | 0.213 | 0.245 | 0.268 | 0.246 | 0.205 | 0.221 | 0.237 | 0.238 |
| 2 | 0.291 | 0.256 | 0.268 | 0.260 | 0.305 | 0.284 | 0.295 | 0.258 | 0.260 | 0.246 |
| 3 | 0.304 | 0.317 | 0.278 | 0.302 | 0.340 | 0.281 | 0.321 | 0.287 | 0.295 | 0.291 |
| 4 | 0.357 | 0.380 | 0.332 | 0.370 | 0.398 | 0.343 | 0.353 | 0.330 | 0.356 | 0.339 |
| 5 | 0.466 | 0.463 | 0.440 | 0.479 | 0.466 | 0.433 | 0.439 | 0.382 | 0.425 | 0.385 |


| Landings weights-at-age (kg) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 0.663 | 0.604 | 0.538 | 0.539 | 0.556 | 0.484 | 0.502 | 0.514 | 0.525 | 0.513 |
| 7 | 0.745 | 0.661 | 0.618 | 0.672 | 0.675 | 0.541 | 0.651 | 0.649 | 0.631 | 0.549 |
| 8 | 0.877 | 0.690 | 0.839 | 0.875 | 0.695 | 0.859 | 0.681 | 0.750 | 0.714 | 0.638 |
| +gp | 1.1007 | 1.1886 | 1.1906 | 1.2018 | 1.0905 | 1.1262 | 1.0389 | 0.9919 | 1.0163 | 0.8369 |
| SOPCOFAC | 1.0002 | 1.0009 | 1.0000 | 1.0007 | 1.0007 | 1.0004 | 0.9994 | 1.0007 | 1.0011 | 1.0008 |
| AGE\YEAR | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| 1 | 0.278 | 0.260 | 0.279 | 0.233 | 0.228 | 0.235 | 0.273 | 0.156 | 0.15 | 0.211 |
| 2 | 0.271 | 0.273 | 0.267 | 0.292 | 0.242 | 0.246 | 0.285 | 0.28 | 0.24 | 0.253 |
| 3 | 0.277 | 0.298 | 0.275 | 0.331 | 0.283 | 0.280 | 0.286 | 0.312 | 0.275 | 0.278 |
| 4 | 0.303 | 0.329 | 0.329 | 0.328 | 0.335 | 0.307 | 0.320 | 0.346 | 0.3 | 0.318 |
| 5 | 0.389 | 0.386 | 0.376 | 0.376 | 0.378 | 0.345 | 0.370 | 0.386 | 0.365 | 0.365 |
| 6 | 0.457 | 0.433 | 0.469 | 0.458 | 0.465 | 0.418 | 0.465 | 0.504 | 0.467 | 0.416 |
| 7 | 0.537 | 0.511 | 0.499 | 0.598 | 0.600 | 0.498 | 0.517 | 0.473 | 0.514 | 0.510 |
| 8 | 0.547 | 0.719 | 0.605 | 0.469 | 0.690 | 0.570 | 0.602 | 0.599 | 0.609 | 0.567 |
| +gp | 0.9862 | 0.9042 | 0.7197 | 1.0433 | 1.1810 | 0.6750 | 0.6550 | 0.735 | 0.946 | 1.003 |
| SOPCOFAC | 1.0005 | 1.0001 | 0.9993 | 1.0002 | 1.0000 | 1.0001 | 0.9994 | 1.001 | 1.002 | 1.005 |
| AGE\YEAR | 2017 | 2018 |  |  |  |  |  |  |  |  |
| 1 | 0.231 | 0.198 |  |  |  |  |  |  |  |  |
| 2 | 0.279 | 0.229 |  |  |  |  |  |  |  |  |
| 3 | 0.289 | 0.262 |  |  |  |  |  |  |  |  |
| 4 | 0.325 | 0.297 |  |  |  |  |  |  |  |  |
| 5 | 0.370 | 0.326 |  |  |  |  |  |  |  |  |
| 6 | 0.426 | 0.407 |  |  |  |  |  |  |  |  |
| 7 | 0.590 | 0.468 |  |  |  |  |  |  |  |  |
| 8 | 0.654 | 0.515 |  |  |  |  |  |  |  |  |
| +gp | 0.7620 | 0.739 |  |  |  |  |  |  |  |  |
| SOPCOFAC | 1.0400 | 0.978 |  |  |  |  |  |  |  |  |

Table 28.7. Plaice in divisions 7.f-g. Discards numbers-at-age.

| Discard numbers-at-age Numbers*10**-3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE\YEAR | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| 1 | 0 | 0 | 0 | 455 | 572 | 542 | 1829 | 73 | 671 | 385 | 960 | 142 | 614 | 83 |
| 2 | 0 | 0 | 0 | 360 | 1211 | 2584 | 3331 | 3595 | 985 | 2719 | 2656 | 2496 | 1283 | 987 |
| 3 | 0 | 0 | 0 | 641 | 441 | 750 | 3408 | 632 | 2041 | 1017 | 1429 | 1950 | 3581 | 1672 |
| 4 | 0 | 0 | 0 | 171 | 118 | 74 | 814 | 393 | 761 | 550 | 1019 | 502 | 1004 | 3195 |
| 5 | 0 | 0 | 0 | 68 | 41 | 47 | 81 | 69 | 399 | 345 | 501 | 179 | 231 | 454 |
| 6 | 0 | 0 | 0 | 3 | 12 | 12 | 32 | 4 | 44 | 54 | 45 | 163 | 32 | 173 |
| 7 | 0 | 0 | 0 | 4 | 4 | 1 | 11 | 1 | 4 | 8 | 99 | 58 | 44 | 77 |
| 8 | 0 | 0 | 0 | 1 | 22 | 1 | 9 | 1 | 5 | 0 | 56 | 25 | 11 | 27 |
| +gp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 8 | 36 |
| TOTALNUM | 0 | 0 | 0 | 1703 | 2421 | 4011 | 9515 | 4768 | 4910 | 5078 | 6765 | 5523 | 6808 | 6704 |
| TONSLAND | 0 | 0 | 0 | 274 | 321 | 453 | 1288 | 583 | 608 | 670 | 1107 | 852 | 1260 | 1158 |
| SOPCOFAC | 0 | 0 | 0 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| AGE\YEAR | 2015 | 2016 | 2017 | 2018 |  |  |  |  |  |  |  |  |  |  |
| 0 | - | - | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 1 | 38 | 29 | 169 | 3 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1527 | 224 | 739 | 92 |  |  |  |  |  |  |  |  |  |  |
| 3 | 1253 | 1610 | 1078 | 587 |  |  |  |  |  |  |  |  |  |  |
| 4 | 753 | 615 | 1257 | 444 |  |  |  |  |  |  |  |  |  |  |
| 5 | 1106 | 229 | 478 | 668 |  |  |  |  |  |  |  |  |  |  |
| 6 | 303 | 209 | 312 | 346 |  |  |  |  |  |  |  |  |  |  |
| 7 | 54 | 34 | 147 | 307 |  |  |  |  |  |  |  |  |  |  |
| 8 | 33 | 15 | 32 | 11 |  |  |  |  |  |  |  |  |  |  |
| +gp | 80 | 9 | 13 | 103 |  |  |  |  |  |  |  |  |  |  |
| TOTALNUM | 5145 | 2974 | 4225 | 2561 |  |  |  |  |  |  |  |  |  |  |
| TONSLAND | 870 | 591 | 895 | 508 |  |  |  |  |  |  |  |  |  |  |
| SOPCOF \% | 1.03 | 1.03 | 0.99 | 1.00 |  |  |  |  |  |  |  |  |  |  |

Table 28.8. Plaice in divisions 7.f-g. Discards weights-at-age.
Discard weights-at-age (kg)

| AGE\YEAR | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0 | 0 | 0 | 0.123 | 0.095 | 0.064 | 0.088 | 0.092 | 0.088 | 0.085 | 0.118 | 0.104 | 0.097 | 0.040 |
| 2 | 0 | 0 | 0 | 0.152 | 0.127 | 0.107 | 0.126 | 0.11 | 0.127 | 0.125 | 0.148 | 0.124 | 0.129 | 0.112 |
| 3 | 0 | 0 | 0 | 0.177 | 0.154 | 0.154 | 0.159 | 0.154 | 0.127 | 0.143 | 0.173 | 0.167 | 0.180 | 0.160 |
| 4 | 0 | 0 | 0 | 0.194 | 0.188 | 0.176 | 0.163 | 0.172 | 0.127 | 0.149 | 0.168 | 0.192 | 0.233 | 0.181 |
| 5 | 0 | 0 | 0 | 0.212 | 0.202 | 0.201 | 0.204 | 0.211 | 0.143 | 0.163 | 0.225 | 0.239 | 0.277 | 0.214 |
| 6 | 0 | 0 | 0 | 0.337 | 0.344 | 0.242 | 0.249 | 0.282 | 0.194 | 0.189 | 0.304 | 0.247 | 0.459 | 0.227 |
| 7 | 0 | 0 | 0 | 0.23 | 0.403 | 0.395 | 0.368 | 0.365 | 0.2 | 0.445 | 0.339 | 0.238 | 0.380 | 0.300 |
| 8 | 0 | 0 | 0 | 0.455 | 0.419 | 0.349 | 0.425 | 0.283 | 0.257 | 0.523 | 0.389 | 0.337 | 0.312 | 0.470 |
| +gp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.713 |
| Discard weights-at-age (kg) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| AGE\YEAR | 2015 | 2016 | 2017 | 2018 |  |  |  |  |  |  |  |  |  |  |
| 0 |  |  | 0.058 |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.12 | 0.148 | 0.14 | 0.105 |  |  |  |  |  |  |  |  |  |  |
| 2 | 0.124 | 0.153 | 0.147 | 0.126 |  |  |  |  |  |  |  |  |  |  |
| 3 | 0.143 | 0.177 | 0.186 | 0.150 |  |  |  |  |  |  |  |  |  |  |
| 4 | 0.171 | 0.205 | 0.225 | 0.188 |  |  |  |  |  |  |  |  |  |  |
| 5 | 0.219 | 0.261 | 0.258 | 0.182 |  |  |  |  |  |  |  |  |  |  |
| 6 | 0.315 | 0.288 | 0.324 | 0.207 |  |  |  |  |  |  |  |  |  |  |
| 7 | 0.208 | 0.341 | 0.271 | 0.324 |  |  |  |  |  |  |  |  |  |  |
| 8 | 0.204 | 0.416 | 0.29 | 0.350 |  |  |  |  |  |  |  |  |  |  |
| +gp | 0.529 | 0.462 | 0.442 | 0.873 |  |  |  |  |  |  |  |  |  |  |

Table 28.9. Plaice in divisions 7.f-g. Stock weights-at-age.

| Stock weights-at-age (kg) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE\YEAR | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 0.112 | 0.086 | 0.107 | 0.109 | 0.082 | 0.096 | 0.103 | 0.256 | 0.075 | 0.000 |
| 2 | 0.216 | 0.170 | 0.212 | 0.217 | 0.167 | 0.192 | 0.206 | 0.298 | 0.193 | 0.087 |
| 3 | 0.315 | 0.252 | 0.313 | 0.322 | 0.257 | 0.288 | 0.307 | 0.352 | 0.307 | 0.232 |
| 4 | 0.406 | 0.334 | 0.412 | 0.426 | 0.350 | 0.383 | 0.408 | 0.418 | 0.417 | 0.369 |
| 5 | 0.492 | 0.414 | 0.507 | 0.528 | 0.447 | 0.479 | 0.507 | 0.495 | 0.521 | 0.498 |
| 6 | 0.570 | 0.493 | 0.599 | 0.628 | 0.548 | 0.574 | 0.606 | 0.584 | 0.621 | 0.619 |
| 7 | 0.642 | 0.570 | 0.689 | 0.727 | 0.653 | 0.668 | 0.704 | 0.685 | 0.717 | 0.733 |
| 8 | 0.707 | 0.646 | 0.775 | 0.823 | 0.762 | 0.763 | 0.801 | 0.797 | 0.808 | 0.839 |
| +gp | 0.839 | 0.822 | 1.015 | 1.132 | 1.129 | 1.049 | 1.114 | 1.190 | 0.965 | 1.064 |
| AGE\YEAR | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | 0.089 | 0.249 | 0.066 | 0.228 | 0.173 | 0.092 | 0.171 | 0.236 | 0.219 | 0.000 |
| 2 | 0.168 | 0.273 | 0.139 | 0.254 | 0.229 | 0.203 | 0.263 | 0.296 | 0.254 | 0.247 |
| 3 | 0.248 | 0.305 | 0.215 | 0.288 | 0.293 | 0.310 | 0.314 | 0.308 | 0.304 | 0.295 |
| 4 | 0.328 | 0.346 | 0.295 | 0.332 | 0.363 | 0.414 | 0.405 | 0.397 | 0.364 | 0.349 |
| 5 | 0.408 | 0.395 | 0.380 | 0.386 | 0.440 | 0.514 | 0.500 | 0.455 | 0.485 | 0.512 |
| 6 | 0.489 | 0.453 | 0.468 | 0.448 | 0.523 | 0.611 | 0.598 | 0.598 | 0.603 | 0.553 |
| 7 | 0.571 | 0.518 | 0.560 | 0.520 | 0.613 | 0.705 | 0.643 | 0.801 | 0.714 | 0.523 |
| 8 | 0.653 | 0.593 | 0.657 | 0.602 | 0.710 | 0.795 | 0.728 | 0.728 | 0.752 | 0.947 |
| +gp | 0.822 | 0.837 | 0.938 | 0.854 | 0.987 | 1.000 | 0.989 | 0.959 | 1.066 | 1.067 |
| AGE\YEAR | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 0.249 | 0.213 | 0.213 | 0.245 | 0.268 | 0.246 | 0.205 | 0.221 | 0.237 | 0.238 |
| 2 | 0.291 | 0.256 | 0.268 | 0.260 | 0.305 | 0.284 | 0.295 | 0.258 | 0.260 | 0.246 |
| 3 | 0.304 | 0.317 | 0.278 | 0.302 | 0.340 | 0.281 | 0.321 | 0.287 | 0.295 | 0.291 |
| 4 | 0.357 | 0.380 | 0.332 | 0.370 | 0.398 | 0.343 | 0.353 | 0.330 | 0.356 | 0.339 |
| 5 | 0.466 | 0.463 | 0.440 | 0.479 | 0.466 | 0.433 | 0.439 | 0.382 | 0.425 | 0.385 |
| 6 | 0.663 | 0.604 | 0.538 | 0.539 | 0.556 | 0.484 | 0.502 | 0.514 | 0.525 | 0.513 |
| 7 | 0.745 | 0.661 | 0.618 | 0.672 | 0.675 | 0.541 | 0.651 | 0.649 | 0.631 | 0.549 |

Stock weights-at-age (kg)

| 8 | 0.877 | 0.690 | 0.839 | 0.875 | 0.695 | 0.859 | 0.681 | 0.750 | 0.714 | 0.638 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| +gp | 1.101 | 1.189 | 1.191 | 1.202 | 1.091 | 1.126 | 1.039 | 0.992 | 1.016 | 0.837 |
| AGE\YEAR | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| 1 | 0.278 | 0.260 | 0.279 | 0.233 | 0.228 | 0.106 | 0.098 | 0.095 | 0.129 | 0.153 |
| 2 | 0.271 | 0.273 | 0.267 | 0.292 | 0.242 | 0.129 | 0.136 | 0.116 | 0.128 | 0.161 |
| 3 | 0.277 | 0.298 | 0.275 | 0.331 | 0.283 | 0.190 | 0.188 | 0.171 | 0.155 | 0.194 |
| 4 | 0.303 | 0.329 | 0.329 | 0.328 | 0.335 | 0.234 | 0.257 | 0.202 | 0.202 | 0.233 |
| 5 | 0.389 | 0.386 | 0.376 | 0.376 | 0.378 | 0.290 | 0.319 | 0.275 | 0.259 | 0.307 |
| 6 | 0.457 | 0.433 | 0.469 | 0.458 | 0.465 | 0.332 | 0.463 | 0.334 | 0.36 | 0.355 |
| 7 | 0.537 | 0.511 | 0.499 | 0.598 | 0.600 | 0.375 | 0.465 | 0.353 | 0.343 | 0.465 |
| 8 | 0.547 | 0.719 | 0.605 | 0.469 | 0.690 | 0.470 | 0.525 | 0.543 | 0.339 | 0.527 |
| +gp | 0.986 | 0.904 | 0.720 | 1.043 | 1.181 | 0.549 | 0.654 | 0.594 | 0.563 | 0.998 |
| AGE\YEAR | 2017 | 2018 |  |  |  |  |  |  |  |  |
| 0 | 0.058 |  |  |  |  |  |  |  |  |  |
| 1 | 0.14 | 0.150 |  |  |  |  |  |  |  |  |
| 2 | 0.153 | 0.152 |  |  |  |  |  |  |  |  |
| 3 | 0.191 | 0.172 |  |  |  |  |  |  |  |  |
| 4 | 0.248 | 0.212 |  |  |  |  |  |  |  |  |
| 5 | 0.286 | 0.235 |  |  |  |  |  |  |  |  |
| 6 | 0.350 | 0.270 |  |  |  |  |  |  |  |  |
| 7 | 0.429 | 0.357 |  |  |  |  |  |  |  |  |
| 8 | 0.522 | 0.498 |  |  |  |  |  |  |  |  |
| +gp | 0.675 | 0.838 |  |  |  |  |  |  |  |  |

Table 28.10. Plaice in divisions 7.f-g: Survey abundance indices.

| IRGFS |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 2018 |  |  |  |  |  |  |
| 1 | 1 | 0.79 | 0.92 |  |  |  |  |
| 1 | 7 |  |  |  |  |  |  |
| 1 | 0.0 | 3.2 | 6.0 | 2.7 | 0.6 | 0.2 | 0.1 |
| 1 | 0.1 | 0.4 | 1.9 | 3.1 | 1.2 | 0.8 | 0.1 |
| 1 | 2.8 | 4.4 | 5.9 | 1.3 | 0.7 | 0.2 | 0.2 |
| 1 | 0.2 | 6.0 | 4.6 | 1.2 | 1.0 | 0.6 | 0.7 |
| 1 | 0.1 | 2.6 | 8.5 | 3.5 | 1.1 | 0.5 | 0.4 |
| 1 | 0.4 | 6.0 | 5.6 | 3.8 | 1.0 | 0.4 | 0.2 |
| 1 | 12.5 | 11.7 | 32.3 | 14.6 | 5.9 | 1.2 | 0.9 |
| 1 | 10.1 | 37.9 | 13.2 | 20.8 | 8.6 | 3.7 | 1.0 |
| 1 | 10.8 | 49.5 | 30.2 | 8.4 | 9.1 | 3.6 | 4.6 |
| 1 | 14.6 | 40.5 | 36.8 | 11.3 | 2.1 | 2.0 | 2.9 |
| 1 | 1.5 | 16.1 | 37.3 | 19.7 | 7.2 | 1.9 | 6.2 |
| 1 | 0.4 | 7.9 | 14.3 | 13.6 | 6.1 | 3.4 | 2.2 |
| 1 | 0.8 | 37.8 | 28.2 | 13.0 | 15.2 | 3.0 | 5.0 |
| 1 | 1.1 | 13.8 | 33.6 | 13.9 | 9.2 | 9.0 | 4.2 |
| 1 | 0.8 | 11.5 | 12.8 | 13.0 | 10.8 | 3.7 | 4.6 |
| 1 | 0.1 | 5.5 | 9.8 | 6.6 | 7.9 | 3.2 | 3.2 |


| E+W BT Survey |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 19952018 |  |  |  |  |  |
| 110.750 .85 |  |  |  |  |  |
| 15 |  |  |  |  |  |
| 1 | 239.590 | 90.480 | 17.230 | 2.960 | 6.840 |
| 1 | 223.690 | 288.110 | 30.780 | 0.990 | 2.620 |
| 1 | 225.370 | 102.140 | 34.540 | 4.250 | 1.770 |
| 1 | 237.200 | 126.220 | 46.990 | 8.920 | 2.000 |
| 1 | 152.590 | 79.620 | 29.030 | 19.670 | 7.000 |
| 1 | 339.630 | 63.170 | 31.250 | 6.560 | 5.500 |
| 1 | 211.440 | 156.140 | 15.810 | 8.740 | 4.230 |
| 1 | 136.740 | 175.120 | 80.450 | 5.930 | 6.130 |
| 1 | 98.370 | 80.480 | 60.950 | 21.830 | 2.720 |
| 1 | 258.510 | 33.410 | 27.080 | 13.420 | 2.190 |
| 1 | 192.500 | 75.220 | 20.870 | 8.060 | 10.930 |
| 1 | 85.780 | 101.970 | 34.160 | 9.570 | 1.790 |
| 1 | 150.400 | 92.250 | 47.260 | 15.110 | 1.670 |
| 1 | 140.690 | 217.040 | 46.790 | 15.700 | 4.820 |
| 1 | 161.810 | 55.960 | 78.580 | 21.450 | 10.890 |
| 1 | 331.760 | 88.540 | 26.410 | 39.940 | 6.680 |
| 1 | 362.260 | 300.140 | 55.040 | 21.860 | 21.370 |
| 1 | 142.130 | 430.790 | 100.570 | 22.360 | 9.020 |
| 1 | 329.790 | 139.060 | 185.390 | 46.850 | 5.770 |
| 1 | 371.760 | 202.300 | 64.650 | 105.700 | 23.800 |
| 1 | 28.360 | 454.080 | 162.340 | 52.370 | 76.660 |
| 1 | 12.520 | 163.100 | 268.260 | 102.300 | 27.500 |
| 1 | 11.490 | 104.1 | 137.39 | 121.110 | 91.87 |
| 1 | 4.15 | 45.26 | 90.2 | 58.1 | 75.08 |

Table 28.11. Plaice in divisions 7.f-g: Commercial fleet abundance indices.

| UK (E+W) BEAM TRAWL 7F. |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 19902010 |  |  |  |  |  |
| 1101 |  |  |  |  |  |
| 48 |  |  |  |  |  |
| 1 | 12.6 | 3.656 | 2.103 | 0.868 | 0.725 |
| 1 | 8.372 | 5.158 | 1.715 | 0.894 | 0.834 |
| 1 | 2.254 | 3.289 | 1.93 | 0.528 | 0.162 |
| 1 | 1.528 | 0.947 | 1.498 | 0.923 | 0.443 |
| 1 | 2.245 | 0.424 | 0.415 | 0.347 | 0.446 |
| 1 | 1.715 | 1.289 | 0.43 | 0.252 | 0.278 |
| 1 | 0.569 | 0.569 | 0.535 | 0.159 | 0.184 |
| 1 | 0.909 | 0.319 | 0.256 | 0.169 | 0.026 |
| 1 | 2.221 | 0.618 | 0.127 | 0.151 | 0.095 |
| 1 | 1.72 | 0.844 | 0.252 | 0.078 | 0.062 |
| 1 | 0.858 | 0.568 | 0.405 | 0.156 | 0.057 |
| 1 | 0.867 | 0.558 | 0.318 | 0.186 | 0.076 |
| 1 | 0.637 | 0.294 | 0.279 | 0.143 | 0.079 |
| 1 | 1.349 | 0.393 | 0.199 | 0.135 | 0.094 |
| 1 | 1.051 | 0.711 | 0.136 | 0.104 | 0.08 |
| 1 | 0.671 | 0.396 | 0.269 | 0.102 | 0.061 |
| 1 | 0.353 | 0.338 | 0.233 | 0.12 | 0.03 |
| 1 | 0.853 | 0.227 | 0.142 | 0.099 | 0.043 |
| 1 | 1.506 | 0.433 | 0.158 | 0.117 | 0.075 |
| 1 | 1.375 | 0.968 | 0.271 | 0.09 | 0.054 |
| 1 | 1.601 | 0.62 | 0.508 | 0.146 | 0.009 |


| UK E+W OTTER TRAWL 7F |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 2010 |  |  |  |  |
| 1101 |  |  |  |  |  |
| 48 |  |  |  |  |  |
| 1 | 6.366 | 2.37 | 0.766 | 0.518 | 0.041 |
| 1 | 10.452 | 2.774 | 1.074 | 0.333 | 0.35 |
| 1 | 7.29 | 3.415 | 1.529 | 0.413 | 0.46 |
| 1 | 1.391 | 2.059 | 0.946 | 0.156 | 0.045 |
| 1 | 1.065 | 0.479 | 0.754 | 0.491 | 0.335 |
| 1 | 2.407 | 0.433 | 0.498 | 0.225 | 0.273 |
| 1 | 2.5 | 0.948 | 0.276 | 0.138 | 0.121 |
| 1 | 0.725 | 0.574 | 0.422 | 0.169 | 0.186 |
| 1 | 0.953 | 0.208 | 0.121 | 0.069 | 0.017 |
| 1 | 1.664 | 0.387 | 0.097 | 0.135 | 0.039 |
| 1 | 1.997 | 0.961 | 0.228 | 0.051 | 0.025 |
| 1 | 2.327 | 0.882 | 0.458 | 0.141 | 0.035 |
| 1 | 1.326 | 0.809 | 0.42 | 0.194 | 0.065 |
| 1 | 0.696 | 0.36 | 0.264 | 0.12 | 0.048 |
| 1 | 1.335 | 0.302 | 0.187 | 0.129 | 0.086 |
| 1 | 1.622 | 0.905 | 0.14 | 0.078 | 0.047 |
| 1 | 0.628 | 0.331 | 0.171 | 0.057 | 0.034 |
| 1 | 0.736 | 0.703 | 0.487 | 0.26 | 0.065 |
| 1 | 0.939 | 0.276 | 0.175 | 0.125 | 0.063 |
| 1 | 1.645 | 0.52 | 0.197 | 0.098 | 0.056 |
| 1 | 0.731 | 0.472 | 0.122 | 0.046 | 0.03 |
| 1 | 1.311 | 0.496 | 0.407 | 0.089 | 0.018 |

Table 28.12. Plaice in divisions 7.f-g: Reconstructed annual catches and and abundance indices used for SPiCT modelling.

| Year | Catch | IRGFS | E+W BT Survey | UK (E+W) BEAM TRAWL 7F. |
| :--- | :--- | :--- | :--- | :--- |$\quad$ UK E+W OTTER TRAWL 7F


| Year | Catch | IRGFS | E+W BT Survey | UK (E+W) BEAM TRAWL 7F. | UK E+W OTTER TRAWL 7F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 707 | 8.3 | 44.0 | 0.027 | 1.221 |
| 2006 | 857 | 8.1 | 57.8 | 0.033 | 2.251 |
| 2007 | 1698 | 14.0 | 69.9 | 0.038 | 1.578 |
| 2008 | 1020 | 11.0 | 71.5 | 0.046 | 2.516 |
| 2009 | 1089 | 54.9 | 117.4 | 0.065 | 1.401 |
| 2010 | 1112 | 47.3 | 80.8 | 0.044 | 2.321 |
| 2011 | 1534 | 56.0 | 118.3 |  |  |
| 2012 | 1565 | 55.0 | 150.3 |  |  |
| 2013 | 1688 | 72.3 | 253.9 |  |  |
| 2014 | 1183 | 39.5 | 203.9 |  |  |
| 2015 | 1159 | 64.2 | 371.1 |  |  |
| 2016 | 1002 | 69.9 | 455.6 |  |  |
| 2017 | 1285 | 45.0 | 408.0 |  |  |
| 2018 | 930 | 30.6 | 319 |  |  |

Table 28.13. Diagnostic of the SPiCT model, stochastic and deterministic reference points.
Convergence: 0 MSG: relative convergence (4)
Objective function at optimum: 56.1174464
Euler time step (years): $1 / 16$ or 0.0625
Nobs C: 42, Nobs I1: 24, Nobs I2: 16, Nobs I3: 16, Nobs 14: 16
Residual diagnostics (p-values)
shapiro bias acf LBox shapiro bias acf LBox
C 0.14320 .27250 .11350 .3794
I1 0.20820 .61110 .01060 .0061 - - * **
120.46630 .60750 .02860 .0773 - ${ }^{*}$.
130.50640 .85540 .40110 .6277 - - -
140.88320 .92210 .07420 .2635

Priors
logn ~ dnorm[ $\left.\log (2), 2^{\wedge} 2\right]$
logalpha ~ dnorm $\left[\log (1), 2^{\wedge} 2\right]$
logbeta ~ dnorm $\left[\log (1), 2^{\wedge} 2\right]$
Model parameter estimates w 95\% Cl
estimate cilow ciupp log.est
alpha1 $2.61349111 .20460215 .670201 \mathrm{e}+000.9606869$
alpha2 $2.85212811 .25787126 .466985 \mathrm{e}+001.0480654$
alpha3 $1.51932370 .60368483 .823758 \mathrm{e}+00 \quad 0.4182653$
alpha4 $2.6261470 \quad 1.19644755 .764272 \mathrm{e}+00 \quad 0.9655178$
beta $0.5195236 \quad 0.21881411 .233489 \mathrm{e}+00-0.6548431$
$r \quad 1.11601690 .39151143 .181245 \mathrm{e}+000.1097660$
rc $0.92084310 .53493461 .585151 e+00-0.0824656$
rold $0.7837734 \quad 0.36383391 .688410 \mathrm{e}+00-0.2436353$
m 2288.7388242 1888.8760912 2.773250e+03 7.7357562
K $\quad 9257.28794605082 .34143681 .686179 \mathrm{e}+049.1331664$
q1 $\quad 0.0341978 \quad 0.01830756 .388030 \mathrm{e}-02-3.3755948$
q2 $0.0079775 \quad 0.00421281 .510610 \mathrm{e}-02-4.8311364$
q3 $0.00002650 .00001385 .080000 \mathrm{e}-05-10.5383081$
q4 $\quad 0.0012099 \quad 0.00061852 .367000 \mathrm{e}-03-6.7171884$
$\begin{array}{llllll}\mathrm{n} & 2.4239025 & 1.03928695 .653206 \mathrm{e}+00 & 0.8853788\end{array}$
sdb $0.15775390 .07919423 .142438 \mathrm{e}-01-1.8467193$
sdf $0.1847780 \quad 0.11933962 .860986 \mathrm{e}-01-1.6886004$
sdi1 $0.4122883 \quad 0.29525525 .757110 \mathrm{e}-01-0.8860324$
sdi2 $0.4499342 \quad 0.30200116 .703314 \mathrm{e}-01-0.7986539$
sdi3 $0.23967920 .14780283 .886673 \mathrm{e}-01-1.4284540$
sdi4 $0.4142848 \quad 0.28545856 .012499 e-01-0.8812016$
sdc $0.0959965 \quad 0.05157381 .786824 \mathrm{e}-01-2.3434435$
Deterministic reference points (Drp)
estimate cilow ciupp log.est
Bmsyd 4970.96379592846 .88710428679 .82471938 .5113690
Fmsyd $0.46042150 .26746730 .7925753-0.7756128$
MSYd 2288.73882421888 .87609122773 .24988647 .7357562
Stochastic reference points (Srp)
estimate cilow ciupp log.est
Bmsys 4846.54665582777 .73813058456 .16591018 .4860217

```
Fmsys 0.4523549 0.2602123 0.7863771-0.7932882
MSYs 2191.3556555 1805.5030743 2659.6684752 7.6922757
    rel.diff.Drp
Bmsys -0.02567130
Fmsys -0.01783255
MSYs -0.04443969
States w 95% Cl (inp$msytype: s)
            estimate cilow ciupp
B_2019.00 7514.5969175 4177.7335503 1.351670e+04
F_2019.00 0.1318512 0.0711697 2.442717e-01
B_2019.00/Bmsy 1.5505054 1.1940580 2.013359e+00
F_2019.00/Fmsy 0.2914774 0.2020672 4.204494e-01
    log.est
B_2019.00 8.924603
F_2019.00 -2.026081
B_2019.00/Bmsy 0.438581
F_2019.00/Fmsy-1.232793
Predictions w 95% Cl (inp$msytype: s)
    prediction cilow ciupp
B_2020.00 7688.8436736 4289.3439279 1.378260e+04
F_2020.00 0.1246514 0.0653057 2.379268e-01
B_2020.00/Bmsy 1.5864582 1.2025618 2.092907e+00
F_2020.00/Fmsy 0.2755610 0.1814337 4.185213e-01
Catch_2020.00 974.4813375 663.22760271.431807e+03
E(B_inf) 7909.1638735 NA NA
            log.est
B_2020.00 8.947526
F_2020.00 -2.082234
B_2020.00/Bmsy 0.461504
F_2020.00/Fmsy-1.288946
Catch_2020.00 6.881905
E(B_inf) 8.975777
```

Table 28.14. Output of the SPiCT model: B (biomass), F (Fishing mortality), B/ Bmš and F/ Fmsy. Estimates (est), upper (upp) and lower (low) $95 \%$ CI. Weights are in tonnes.

|  |  |  | $\begin{aligned} & \frac{0}{2} \\ & \frac{1}{2} \\ & \frac{n}{4} \\ & \frac{14}{2} \end{aligned}$ |  |  | $\begin{aligned} & \text { on } \\ & \frac{1}{3} \\ & \frac{1}{n} \\ & \stackrel{\infty}{\infty} \end{aligned}$ | $\begin{aligned} & 3 \\ & \frac{0}{\infty} \end{aligned}$ | $\frac{\overline{\tilde{y}}}{\stackrel{\sim}{\infty}}$ | $\stackrel{\text { O}}{\text { 을 }}$ | $\begin{aligned} & 3 \\ & \frac{0}{4} \end{aligned}$ | $\begin{aligned} & \overline{\tilde{y}} \\ & \frac{6}{4} \end{aligned}$ | $\stackrel{\text { 을 }}{ }$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 0.492 | 0.990 | 1.992 | 0.239 | 0.488 | 1.000 | 899 | 2367 | 6236 | 0.165 | 0.448 | 1.217 |
| 1978 | 0.495 | 0.979 | 1.936 | 0.268 | 0.590 | 1.298 | 1068 | 2860 | 7662 | 0.167 | 0.443 | 1.177 |
| 1979 | 0.448 | 0.912 | 1.857 | 0.300 | 0.672 | 1.510 | 1225 | 3259 | 8670 | 0.156 | 0.412 | 1.092 |
| 1980 | 0.471 | 0.943 | 1.886 | 0.373 | 0.821 | 1.806 | 1574 | 3980 | 10060 | 0.169 | 0.426 | 1.074 |
| 1981 | 0.524 | 1.042 | 2.070 | 0.429 | 0.932 | 2.026 | 1836 | 4519 | 11121 | 0.192 | 0.471 | 1.158 |
| 1982 | 0.507 | 1.009 | 2.007 | 0.434 | 0.934 | 2.009 | 1862 | 4525 | 10993 | 0.188 | 0.456 | 1.110 |
| 1983 | 0.473 | 0.927 | 1.816 | 0.441 | 0.925 | 1.941 | 1897 | 4483 | 10598 | 0.177 | 0.419 | 0.991 |
| 1984 | 0.462 | 0.863 | 1.612 | 0.488 | 0.962 | 1.896 | 2102 | 4661 | 10336 | 0.176 | 0.390 | 0.867 |
| 1985 | 0.558 | 0.944 | 1.597 | 0.618 | 1.090 | 1.923 | 2650 | 5285 | 10541 | 0.214 | 0.427 | 0.852 |
| 1986 | 0.676 | 1.059 | 1.658 | 0.702 | 1.134 | 1.832 | 2944 | 5497 | 10262 | 0.257 | 0.479 | 0.892 |
| 1987 | 0.757 | 1.125 | 1.672 | 0.730 | 1.111 | 1.690 | 3007 | 5384 | 9640 | 0.286 | 0.509 | 0.906 |
| 1988 | 0.895 | 1.277 | 1.820 | 0.748 | 1.092 | 1.595 | 3049 | 5292 | 9186 | 0.337 | 0.578 | 0.990 |
| 1989 | 1.049 | 1.462 | 2.039 | 0.706 | 1.012 | 1.450 | 2856 | 4902 | 8415 | 0.392 | 0.662 | 1.117 |
| 1990 | 1.200 | 1.673 | 2.331 | 0.611 | 0.882 | 1.272 | 2458 | 4273 | 7427 | 0.442 | 0.757 | 1.295 |
| 1991 | 1.248 | 1.753 | 2.463 | 0.480 | 0.696 | 1.009 | 1876 | 3372 | 6063 | 0.445 | 0.793 | 1.414 |
| 1992 | 1.171 | 1.644 | 2.307 | 0.390 | 0.562 | 0.809 | 1477 | 2723 | 5020 | 0.407 | 0.744 | 1.359 |
| 1993 | 1.115 | 1.575 | 2.227 | 0.355 | 0.509 | 0.730 | 1329 | 2468 | 4584 | 0.386 | 0.713 | 1.316 |
| 1994 | 1.103 | 1.569 | 2.232 | 0.340 | 0.487 | 0.697 | 1265 | 2358 | 4399 | 0.382 | 0.710 | 1.319 |
| 1995 | 1.101 | 1.574 | 2.248 | 0.329 | 0.468 | 0.666 | 1210 | 2268 | 4252 | 0.380 | 0.712 | 1.333 |
| 1996 | 1.075 | 1.537 | 2.199 | 0.314 | 0.445 | 0.633 | 1145 | 2159 | 4070 | 0.372 | 0.695 | 1.299 |
| 1997 | 1.132 | 1.618 | 2.313 | 0.327 | 0.464 | 0.657 | 1213 | 2247 | 4162 | 0.399 | 0.732 | 1.344 |
| 1998 | 1.203 | 1.726 | 2.477 | 0.318 | 0.456 | 0.655 | 1179 | 2210 | 4143 | 0.421 | 0.781 | 1.448 |
| 1999 | 1.200 | 1.733 | 2.502 | 0.288 | 0.417 | 0.603 | 1056 | 2019 | 3862 | 0.409 | 0.784 | 1.501 |
| 2000 | 1.142 | 1.679 | 2.469 | 0.241 | 0.349 | 0.506 | 855 | 1693 | 3352 | 0.381 | 0.760 | 1.516 |
| 2001 | 1.078 | 1.584 | 2.329 | 0.219 | 0.318 | 0.459 | 769 | 1539 | 3079 | 0.357 | 0.717 | 1.436 |
| 2002 | 1.042 | 1.544 | 2.288 | 0.211 | 0.307 | 0.448 | 737 | 1489 | 3009 | 0.343 | 0.698 | 1.422 |


|  | $\begin{aligned} & 3 \\ & \text { o } \\ & \vdots \\ & \sum_{4}^{n} \\ & \frac{4}{4} \end{aligned}$ |  | $\begin{aligned} & \text { 을 } \\ & \frac{\imath}{n} \\ & \stackrel{1}{2} \\ & \end{aligned}$ |  | $\begin{aligned} & \underset{\tilde{y}}{\stackrel{y}{0}} \\ & \text { 玄 } \\ & \underset{\sim}{\infty} \end{aligned}$ |  | $\begin{aligned} & \frac{3}{\infty} \\ & \frac{0}{\infty} \end{aligned}$ | $\frac{\overline{\bar{U}}}{\stackrel{\bar{W}}{\infty}}$ | $\begin{aligned} & \frac{0}{2} \\ & \frac{0}{7} \\ & \infty \end{aligned}$ | $\begin{aligned} & 3 \\ & \frac{3}{4} \end{aligned}$ | $\frac{\bar{\psi}}{\underline{y}}$ | $\stackrel{\text { 을 }}{ }$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0.976 | 1.456 | 2.171 | 0.206 | 0.303 | 0.446 | 720 | 1470 | 3002 | 0.318 | 0.658 | 1.365 |
| 2004 | 0.925 | 1.398 | 2.114 | 0.196 | 0.286 | 0.419 | 675 | 1388 | 2852 | 0.297 | 0.633 | 1.346 |
| 2005 | 0.880 | 1.324 | 1.992 | 0.181 | 0.259 | 0.371 | 609 | 1256 | 2588 | 0.283 | 0.599 | 1.266 |
| 2006 | 0.858 | 1.269 | 1.878 | 0.194 | 0.275 | 0.391 | 661 | 1335 | 2698 | 0.284 | 0.574 | 1.160 |
| 2007 | 1.023 | 1.511 | 2.233 | 0.258 | 0.360 | 0.503 | 904 | 1745 | 3367 | 0.350 | 0.684 | 1.334 |
| 2008 | 1.001 | 1.520 | 2.309 | 0.261 | 0.377 | 0.545 | 904 | 1828 | 3698 | 0.337 | 0.688 | 1.404 |
| 2009 | 0.783 | 1.157 | 1.709 | 0.282 | 0.405 | 0.581 | 966 | 1962 | 3983 | 0.257 | 0.523 | 1.064 |
| 2010 | 0.654 | 0.984 | 1.482 | 0.352 | 0.513 | 0.748 | 1239 | 2487 | 4992 | 0.219 | 0.445 | 0.907 |
| 2011 | 0.566 | 0.859 | 1.304 | 0.464 | 0.686 | 1.016 | 1688 | 3326 | 6555 | 0.197 | 0.389 | 0.768 |
| 2012 | 0.491 | 0.758 | 1.171 | 0.618 | 0.935 | 1.417 | 2339 | 4533 | 8785 | 0.176 | 0.343 | 0.669 |
| 2013 | 0.415 | 0.643 | 0.996 | 0.772 | 1.158 | 1.739 | 2950 | 5614 | 10681 | 0.152 | 0.291 | 0.558 |
| 2014 | 0.331 | 0.503 | 0.762 | 0.890 | 1.281 | 1.842 | 3348 | 6208 | 11510 | 0.121 | 0.227 | 0.428 |
| 2015 | 0.267 | 0.392 | 0.576 | 0.998 | 1.361 | 1.856 | 3652 | 6597 | 11915 | 0.096 | 0.177 | 0.327 |
| 2016 | 0.232 | 0.332 | 0.474 | 1.136 | 1.496 | 1.971 | 4067 | 7251 | 12926 | 0.082 | 0.150 | 0.274 |
| 2017 | 0.233 | 0.324 | 0.449 | 1.253 | 1.630 | 2.120 | 4427 | 7898 | 14091 | 0.081 | 0.146 | 0.265 |
| 2018 | 0.229 | 0.326 | 0.463 | 1.265 | 1.643 | 2.132 | 4427 | 7961 | 14315 | 0.080 | 0.147 | 0.271 |
| 2019 | 0.202 | 0.291 | 0.420 | 1.194 | 1.551 | 2.013 | 4178 | 7515 | 13517 | 0.071 | 0.132 | 0.244 |

Table 28.15. Short-term projection of the SPiCT model, plaice 7.fg for 2020.

| Reference Point | Estimate | Cl 95\% Low | Cl 95\% upp | CV, \% |
| :--- | :---: | :---: | :---: | :---: |
| B/B $\mathrm{B}_{\text {MSY }}$ | 1.586 | 1.203 | 2.093 | 14.2 |
| F/F MSYS | 0.276 | 0.159 | 0.479 | 21.6 |



Figure 28.1. Plaice in divisions 7.f-g: Landing and discards by different metiers when both landings and discards were sampled simultaneously.


Figure 28.2. Plaice in divisions 7.f-g: Contribution of sampled and unsampled landings and discards to final assessment catch numbers-at-age in 2018.


Figure 28.3. Plaice in divisions 7.f-g: Age composition of international landings (blue) and discards (red) from 2004 to 2018.


Figure 28.4. Trends cpues of surveys.

VIlfg plaice: UK(E\&W) LPUE and CPUE


Vllfg plaice: UK(E\&W) Effort


Figure 28.5. Plaice in divisions 7.f-g: UK (E\&W) lpue and effort by fleet.

7 g plaice: Ireland: fleet LPUE


7g plaice: Ireland: Effort


7fg plaice - Belgian Beam trawl


Figure 28.6. Plaice in divisions 7.f-g: Ireland and Belgium: lpue and effort by fleet.



Figure 28.10. Diagnostics of the SPiCT model.


Figure 28.11. Ple 7 f\&g. Dynamics of $B / B_{\text {msy }}$ and $F / F_{\text {msy }}$ from SPiCT run.

# 26 <br> Plaice (Pleuronectes platessa) in divisions 7h-k (Celtic Sea South, southwest of Ireland) 

## Type of assessment in 2019

An update XSA assessment was performed for the $7 . \mathrm{jk}$ component of the landings according to the stock annex. MSY and PA reference points were estimated.

ICES advice applicable to 2019
ICES advises that when the precautionary approach is applied, there should be zero catches in 2019. Discards are known to take place but cannot be quantified; therefore, total catches cannot be calculated.
http://ices.dk/sites/pub/Publication\ Reports/Advice/2018/2018/ple-7h-k.pdf
ICES advice applicable to 2018
ICES advises that when the precautionary approach is applied, there should be zero catches in 2018. Discards are known to take place but cannot be quantified; therefore, total catches cannot be calculated.
http://ices.dk/sites/pub/Publication\ Reports/Advice/2017/2017/ple.27.7h-k.pdf

### 26.1 General

## Stock description and management units

Plaice in 7.h-k is on the southwestern margins of the species distribution. Although the TAC is set for divisions $7 . h, j$ and $k$, the assessment is performed for $7 . j k$ only. This separation in assessment and TAC area is driven by two main reason: firstly, age-disaggregated data were not historically available for $7 . \mathrm{h}$, and secondly, there is a wide geographical distance of several hundred miles, between the inshore 7 j fishery and the 7 h fishery. This distance would suggest that $7 . \mathrm{h}$ stock may constitute a spate stock, and may be a continuation of the plaice caught in the western English Channel (7.e).

The assessed part of the fishery focusses mainly on 7.j where Irish vessels operate on sandy grounds off the southwest of Ireland. Irish VMS and logbook data indicate that the landings of plaice in 7.j occur close to shore, and are part of a mixed fishery. With plaice forming only a small component ( $<5 \%$ ) of the overall landings per trip (Figure 27.1). Landings in $7 . \mathrm{k}$ are minor.
It is the recommendation of this group that plaice in $7 \mathrm{~h}-\mathrm{k}$ be benchmarked to incorporate new datasets, including IBTS survey indices and age-disaggregated data for 7.h.

## Management applicable to 2019 and 2018

TAC table 2019

| Species: <br> Plaice <br> Pleuronectes platessa |  | Zone: | 7h, 7 j and 7 k <br> (PLE/7HJK.) |
| :---: | :---: | :---: | :---: |
| Belgium | $\left.7{ }^{1}\right)$ |  |  |
| France | $14{ }^{(1)}$ |  |  |
| Ireland | 47 (1) |  |  |
| The Netherlands | 27 (1) |  |  |
| United Kingdom | 14 ( ${ }^{1}$ ) |  |  |
| Union | 109 (1) |  |  |
| TAC | 109 (1) |  | Precautionary TAC Article 8 of this Regulation applies Article 13(1) of this Regulation applies |

${ }^{1}$ ) Exclusively for by-catches of plaice in fisheries for other species. No directed fisheries for plaice are permitted under this quota.

TAC table 2018

| Species: $\begin{aligned} & \text { Plaice } \\ & \text { Plaronectes platessa }\end{aligned}$ |  | Zone: | 7h, 7 j and 7 k (PLEE/7HJK.) |
| :---: | :---: | :---: | :---: |
| Belgium | 8 |  |  |
| France | 16 |  |  |
| Ireland | 56 |  |  |
| The Netherlands | 32 |  |  |
| United Kingdom | 16 |  |  |
| Union | 128 |  |  |
| TAC | 128 |  | Precautionary TAC <br> Article 12(1) of this Regulation applies |

*Article 12 refers to the closure of the Porcupine Bank in May and July.

### 26.2 Data

## Landings and discards

The nominal landings are given in Table 27.1. Historic Belgian landings from 7.j are considered to have been area misreported and have been removed from the total landings. The remainder of Section 27 concerns $7 . j k$ only, as this is the area on which the assessment is run.

Discard and retained catch numbers for the Irish 7.j OTB fleet in 2018 are based on nine observer trips. There is currently no reliable time-series of discards-at-age or discard numbers, therefore they are not included in the assessment. Since 2007, the proportion of the 7.j catch that was discarded varies between $10 \%$ and $100 \%$, however the number of trips in some years was very low. Since 2001, the number of trips has ranged from 2-23 per year and with the average proportion of catch being discarded in that period being $38 \%$. Although not included in this assessment, it is important tonote here that discarding is part of this fishery, and is considered to be quitehigh, and on the increase due to limited quota.

## Commercial effort and Ipue

The commercial effort landings and lpue for the Irish otter trawl fleet in 7.j is shown in Figure 27.1b.

## Landings numbers-at-age

Landings numbers-at-age are given in Table 27.3 and Figure 27.3. Figure 27.4 shows a bubble plot of the standardised landings proportions-at-age. There is very little contrast in the numbers-at-age matrix. Figure 27.5 gives the stock weights (which are the same as the landings weights).

## Biological

Natural mortality was assumed to be 0.12 for all ages and the proportion mature for age 4 is assumed to be 0.86 and 1 for all older ages.

## Surveys and commercial tuning fleets

Historically, there is no survey index available for this stock as the Irish IBTS Q4 Groundfish Survey data were considered to be too noisy to be used. A commercial tuning index is used. This index comes from the Irish VMS data, which are linked to logbook landings (see Gerritsen et al., 2011 for details on linking VMS and logbook data). These data were used to identify plaice fishing grounds, which are targeted by OTB vessels (Figure 27.6), and to estimate the effort and landings of the OTB vessels within these fishing grounds (Figure 27.6). Thelpuetrends identified by Gerritsen et al. (2011) in VMS-based, mirrored the lpue of Irish OTB vessels in the whole of 7.j. However, it should be noted that this index is not sensitive to changes in the spatial distribution of the fleet as it assumes that all vessels operating in 7.j are capable of catching plaice. This is not the case, as only vessels close to the shore catch plaice and not those which operate further offshore.

The age composition of the Irish OTB fleet in 7.j is used as a tuning fleet (Table 27.5). Figure 27.7 shows the log standardised numbers-at-age in the tuning index by year and cohort. No year effects are obvious, but cohort tracking is not particularly good either. This is probably results from the lack of contrast in recruitment (see 'Data quality'). Figure 27.8 shows the internal consistency regressions for the tuning fleet.

In years to come, the annual Irish Beam Trawl Ecosystem Survey (IBES) may act as a possible tuning index for this stock, as a number of valid tows occur in the area where the fishery is executed. The first of these surveys took place in 2016 (ICES, 2016c) and was repeated in 2017 and 2018.

## Data quality

The age data for 1995 were considered insufficient and for this year the combined age data for 1993-1996 were used. Sampling is considered sufficient to establish landings numbers-at-age. The lack cohort tracking in the numbers-at-age matrix is most likely due to an absence of very strong or weak cohorts, rather than poor sampling or ageing.

### 26.3 Historical stock development

Target category:3.2.0.
Model used: XSA
Software used: Lowestoft vpa95.exe and FLR with R version 3.5.1 and packages FLXSA_2.6.2, FLAssess_2.6.3,FLEDA_2.5.2,FLCore_2.6.13.

## Exploratory assessment

Several exploratory assessments were carried out by means of a separable VPA and XSA. The initial VPA runs explored the year and age range to be used in the separable and the choices of reference age, final $F$ and $S$. The XSA runs explored the choices of $q$-age, F-shrinkage and the minimum SE threshold. The results of these are available on the ICES SharePoint site of WGCSE under data for this stock.

## Final assessment

The model was applied to landings numbers for ages 4-8+ for the years 1993-2018. The tuning fleet included ages 4-8 for the years 2007-2018.

Model Options:

| Option | Setting |
| :--- | :---: |
| Ages catch dep stock size | None |
| Q plateau | No |
| Taper | 6 |
| F shrinkage SE | 1.0 |
| F shrinkage year range | 5 |
| F shrinkage age range | No |
| Fleet SE threshold | 3 |
| Prior weights | 0.3 |

The diagnostics of the final XSA assessment are given in Table 27.6. Age classes 4 to $8+$ were included in the model. Younger ages were omitted because significant discarding is expected to take place at these ages. Figure 27.9 shows the residuals. There are some year effects but the absolute values are small. Because the landings and the tuning fleet have nearly identical age compositions, the year effects result from the lpue estimate of the tuning fleet. The retrospective analysis shows no consistent retrospective bias, despite some noise the retrospective seems stable. A summary of relative trends in landings, recruitment, SSB and F is given in Table 27.10 and Figure 27.7.
A Mohn's rho analysis was conducted based on the XSA stock assessment results, i.e. the last data year (2018) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (ages 3-9) | recruitment |
| :--- | :--- | :--- | :--- |
| Mohn's rho value | -0.2204965 | 0.1091162 | -0.2668708 |

The SSB and recruitment Mohn's rho values for this assessment are higher than the threshold of $20 \%$ imposed by ICES for 2019 assessments, therefore the current assessment indicates a low consistency.

### 26.4 MSY evaluation

TheMSY reference points used were the same as last year. These reference points wereestimated using based on WKProxy (ICES, 2016a). FMSY reference point of $F=0.25$ was estimated based on $\mathrm{F}_{0.1}$ from a Thompson-Bell yield-per-recruit analysis of the landings numbers-at-age. This is a data-limited approach (which was in line with the ToRs of WKProxy) and the resulting reference point is not directly comparable with the outputs from the XSA (only the landings data are used in the Thompson-Bell approach). In 2016, this working group (ICES, 2016d) recommended that it would be more appropriate to move the stock to Category 2 next year and to apply the WKMSYREF4 (ICES, 2016b) methodology for estimating reference points (ICES, 2012).
An exploratory MSY evaluation was conducted in 2017,following WKMS YREF4 guidelines, and is presented here. Details on this evaluation can be found in the working document in appendix xxx. The stock-recruitment graph (Figure 27.11) suggests recruitment has been impaired for most of the time-series. Because there is no obvious stock-recruitment relationship (it appears to be a recruit-stock relationship), it is difficult to specify an appropriate SR model. The SR estimation was carried out on age $>=4$ as that is the onset of recruitment using: fit <-eqsr_fit_shift(stock, $n s a m p=1000$, models $=c($ "Segreg" $)$, rshift $=4$ ). From this Blim was estimated to be 203.57 (Blim $<-$ median(fit\$sr.sto\$b.b)) and a $\mathrm{B}_{\mathrm{pa}}$ at 282.88 ( $B_{p a}<-B_{p a}\left(B_{l i m}, 0.2\right)$. The following settings were used to estimate the MSY reference points using the eqsim_run\{msy\} function in the msy package in $R$ (full code available on SharePoint):
stockSetup <-list(data = stock,
bio.years $=c(2007,2016)$,
bio.const = FALSE,
sel. years $=c(2007,2016)$,
sel.const $=$ FALSE,
Fscan $=\operatorname{seq}(0,0.8, b y=0.005)$,
$\mathrm{Fcv}=0.212$,
Fphi $=0.423$,
Blim = Blim,
$\mathrm{Bpa}=\mathrm{Bpa}$,
verbose $=$ TRUE,
extreme.trim $=c(0.05,0.95))$

Where $\mathrm{F}_{\mathrm{cv}}$ and $\mathrm{F}_{\text {phi }}$ were the same as those used by WKMSYREF4 for plaice in 7e (ICES, 2016b), which was calculate during WKMSYREF3 (ICES, 2014). Figures 7.11.12 and 7.11.13 summarise the MSY evaluation. The analysis resulted in an estimate of $\mathrm{F}_{\text {MSY }}=0.289$ without a $\mathrm{B}_{\text {trigger }}$ harvest control rule and $\mathrm{F}_{\mathrm{MSY}}=0.306$ with a $\mathrm{B}_{\text {trigger }}=\mathrm{B}_{\text {pa }} \mathrm{HCR}$. These values are slightly higher than the Fisy proxy of 0.25 proposed by WKProxy (ICES, 2016a). $_{\text {a }}$

Biological reference points

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 282 | $\mathrm{B}_{\mathrm{pa}}$ | ICES (2017) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.289 | Median point estimates of Eqsim with segmented regression S-R relationship | ICES <br> (2017) |
| Precautionary approach | $\mathrm{Blim}_{\text {lim }}$ | 203 | Break point segmented regression S-R relationship | ICES (2017) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 282 | $\mathrm{B}_{\text {lim }} \times \operatorname{esp}(1.645 \times \sigma) ; \sigma=0.20$ | ICES (2017) |
|  | $\mathrm{F}_{\text {lim }}$ | 0.471 | F with $50 \%$ probability of SSB $<\mathrm{Bl}_{\text {lim }}$ | ICES (2017) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.339 | $\mathrm{F}_{\text {lim }} \times \exp (-1.645 \times \sigma) ; \sigma=0.20$ | ICES <br> (2017) |
| Management plan | SSB ${ }_{\text {mgt }}$ | Not applicable |  |  |
|  | $\mathrm{F}_{\text {mgt }}$ | Not applicable |  |  |

### 26.5 Uncertainties and bias in the assessment and forecast

The advice is based on an assessment model accepted for trends, used as an indicator of stock size. The uncertainty associated with the index values is not available. The assessment is only based on ages 4 and older; ICES does not have reliable information on younger ages.

The assessment is carried out on the landings in divisions 7.j and 7.k. The trends in this area may not be representative of the whole stock area (7.hjk). Historically, no age information was available for Division 7.h, therefore ICES was unable to assess stock trends in Division 7.h. However, age data for 7.h were submitted to InterCatch in 2017, 2018 and 2019, therefore a benchmark should be considered for this stock. The advice takes into account the reported landings from the full TAC area; divisions 7.h-k.

The apparent reduction in SSB in 2015 is mainly driven by a reduction in relative abundance of young fish in recent years, there is a slight increase in 2016, but is again showing a downward trend in 2018. It is unclear, whether this lack of young fish in the landings (and commercial tuning lpue index) is due to increased discarding or poor recruitment (Table 27.1). There has been an overall decrease in landings, with an increase in landings by Ireland and a reduction in those by France and the UK.

The tuning index only begins in 2006 and there is limited contrast between the cohorts; therefore the assessment is driven mostly by the strong trend in 7.jk landings in the first ten years of the time-series.

Discards in this stock may be considerable but are not presently included in the model as there are insufficient data, and because this might introduce more noise in the catch numbers-at-age matrix, particularly in the early years of the time-series when sampling levels were variable.

The use of a commercial tuning fleet has the potential to introduce bias if the behaviour of the fleet changes; for example, the spatial distribution of effort can change over time, resulting in higher or lower catch rates of certain species. Additionally, changes to the gear, vessel power, towing speed, etc. can influence the catch rates. By limiting the index to an area where plaice are known to be caught, some of this potential bias will be avoided. The working group applied a spatial stratification to check that changes in effort distribution within the plaice area did not affect the index and this did not appear to be the case. Because the stratified estimate is likely to be less precise, the final tuning index was based on the un-stratified estimate. More sophisticated modelling approaches to standardise the commercial index could be investigated for a future benchmark.

### 26.6 Recommendations for the next benchmark

This stock is scheduled to be benchmarked in 2021.

### 26.7 Management considerations

Plaice are taken as a minor bycatch in a mixed fishery and should be managed as such. Restricting the landings by TAC is unlikely to reduce the catches. As per ICES guidelines the PA buffer was applied in 2019.
Because plaice are caught in spatially distinct areas, restricting effort in these areas will be more effective than limiting landings. Additionally, management should focus on reducing discards. The recently introduced square mesh panels will be unlikely to effect on catches of undersized plaice. An increase in mesh size could improve selection, but will also affect the catches of marketable fish. The landings obligation is not currently in effect for this stock.

The TAC area includes Division 7.h. However, the landings from divisions 7.jk are taken in the northeastern part of Division 7.j, which is remote from the northern part of Division 7.h, where most of the Division 7.h landings are taken. It is likely that the plaice from Division 7.h are part of the divisions 7.e or 7.fg stocks. Although no further information on stock structure is likely to become available, this stock should be benchmarked.

Historically, only landings data were available for Division 7.h, however age-structured data were available in InterCatch in 2017, 2018 and 2019 for this area, therefore a benchmark should take place to determine how best to incorporate these samples. Landings in Division 7.h have fluctuated around $50 \%$ of the total landings of the stock (i.e. in divisions 7.h-k) since 1993.

### 26.8 References

Gerritsen HD and Lordan C. 2011. Integrating Vessel Monitoring Systems (VMS) data with daily catch data from logbooks to explore the spatial distribution of catch and effort at high resolution. ICES J Mar Sci 68 (1): 245-252.

ICES. 2012. ICES implementation of advice for data-limited stocks in 2012. Report in support of ICES advice. ICES CM 2012/ACOM:68.

ICES. 2014. Report of the Joint ICES-MYFISH Workshop to consider the basis for Fmsy ranges for all stocks (WKMS YREF3), 17-21 November 2014, Charlottenlund, Denmark. ICES CM 2014/ACOM:64. 164 pp.

ICES. 2016a. Report of the Workshop to consider MSY proxies for stocks in ICES category 3 and 4 stocks in Western Waters (WKProxy), 3-6 November 2015, ICES Headquarters, Copenhagen. ICES CM 2015/ACOM:61. 183 pp .

ICES. 2016b. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

ICES. 2016c. Final Report of the Working Group on Beam Trawl Surveys (WGBEAM), 12-15 April 2016, La Rochelle, France. ICES CM 2016/ SSGIEOM:20.125 pp.
ICES. 2016d. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 4-13 May 2016, Copenhagen, Denmark. ICES CM 2016ACOM:13. 1312 pp.

Table 27.1. Plaice in divisions 7. h-k (Southwest Ireland). Nominal landings ( t ), 1993-2018, as officially reported to ICES.

|  | 7.jk |  |  |  |  | 7.h |  |  |  |  | 7.jk | 7.h | 7.hjk | 7.hjk |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | BEL | FRA | IRL | UK | OTH | BEL | FRA | IRL | UK | OTH | TOT* | TOT | TOT | WG est |
| 1993 | 0 | 8 | 383 | 46 | 0 | 0 | 56 | 0 | 179 | 0 | 437 | 235 | 672 | 655 |
| 1994 | 0 | 6 | 251 | 60 | 0 | 0 | 42 | 20 | 199 | 0 | 317 | 261 | 578 | 577 |
| 1995 | 0 | 12 | 317 | 90 | 0 | 0 | 48 | 4 | 196 | 0 | 419 | 248 | 667 | 542 |
| 1996 | 0 | 3 | 295 | 38 | 0 | 0 | 45 | 10 | 117 | 52 | 336 | 224 | 560 | 453 |
| 1997 | 0 | 6 | 337 | 32 | 0 | 0 | 63 | 7 | 106 | 0 | 375 | 176 | 551 | 645 |
| 1998 | 0 | 8 | 282 | 16 | 0 | 0 | 41 | 4 | 90 | 13 | 306 | 148 | 454 | 444 |
| 1999 | 42 | 0 | 296 | 15 | 0 | 3 | 0 | 3 | 67 | 1 | 311 | 74 | 385 | 406 |
| 2000 | 4 | 16 | 195 | 9 | 5 | 0 | 38 | 5 | 67 | 2 | 225 | 112 | 337 | 299 |
| 2001 | 0 | 16 | 157 | 6 | 3 | 27 | 34 | 3 | 67 | 0 | 182 | 131 | 313 | 261 |
| 2002 | 14 | 21 | 155 | 5 | 2 | 55 | 24 | 0 | 54 | 0 | 183 | 133 | 316 | 313 |
| 2003 | 4 | 7 | 125 | 9 | 6 | 16 | 25 | 2 | 47 | 0 | 147 | 90 | 237 | 217 |
| 2004 | 0 | 5 | 87 | 6 | 6 | 67 | 27 | 4 | 30 | 0 | 104 | 128 | 232 | 221 |
| 2005 | 0 | 4 | 88 | 2 | 0 | 32 | 16 | 2 | 26 | 0 | 94 | 76 | 170 | 164 |
| 2006 | 1 | 6 | 63 | 1 | 1 | 22 | 31 | 2 | 17 | 0 | 71 | 72 | 143 | 147 |
| 2007 | 2 | 9 | 72 | 2 | 11 | 7 | 21 | 0 | 18 | 2 | 94 | 48 | 142 | 120 |
| 2008 | 3 | 5 | 72 | 1 | 1 | 25 | 7 | 0 | 11 | 0 | 79 | 43 | 122 | 135 |


|  | 7.jk |  |  |  |  | 7.h |  |  |  |  | 7.jk | 7.h | 7.hjk | 7.hjk |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | BEL | FRA | IRL | UK | OTH | BEL | FRA | IRL | UK | OTH | TOT* | TOT | TOT | WG est |
| 2009 | 4 | 7 | 71 | 2 | 0 | 1 | 37 | 0 | 30 | 0 | 80 | 68 | 148 | 148 |
| 2010 | 5 | 11 | 66 | 1 | 0 | 0 | 44 | 0 | 34 | 0 | 78 | 78 | 156 | 155 |
| 2011 | 6 | 11 | 67 | 2 | 0 | 4 | 47 | 6 | 42 | 0 | 80 | 99 | 179 | 178 |
| 2012 | 7 | 17 | 93 | 0 | 0 | 2 | 45 | 6 | 36 | 0 | 110 | 89 | 199 | 196 |
| 2013 | 0 | 14 | 51 | 0 | 0 | 0 | 35 | 1 | 40 | 0 | 65 | 76 | 141 | 182 |
| 2014 | 0 | 11 | 74 | 0 | 0 | 4 | 40 | 4 | 15 | 0 | 85 | 63 | 148 | 169 |
| 2015 | 0 | 10 | 23 | 0 | 0 | 5 | 50 | 2 | 17 | 0 | 33 | 73 | 107 | 114 |
| 2016 | 0 | 7 | 30 | 0 | 0 | 7 | 39 | 2 | 15 | 0 | 37 | 63 | 100 | 99 |
| 2017 | 0 | 12 | 39 | 1 | 0 | 11 | 41 | 3 | 9 | 0 | 52 | 64 | 116 | 115 |
| 2018* | 1 | 5 | 30 | 0 | 0 | 16 | 30 | 2 | 0 | 11 | 59 | 36 | 95 | 97 |

* Excluding Belgium

Table 27.3. Landings numbers-at-age for plaice in 7.jk, 1993-2018.

|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 92.8 | 623.6 | 479.4 | 115.4 | 44.8 | 22.8 | 10.5 | 5.9 | 2.6 |
| 1994 | 103.7 | 340.2 | 259.7 | 82.1 | 45.5 | 18.3 | 8.1 | 5.0 | 2.9 |
| 1995 | 207.3 | 632.8 | 347.5 | 106.9 | 36.3 | 15.7 | 7.1 | 4.8 | 3.1 |
| 1996 | 76.9 | 314.5 | 228.1 | 127.0 | 37.1 | 23.4 | 4.9 | 3.0 | 0.7 |
| 1997 | 166.4 | 277.0 | 268.1 | 118.9 | 42.3 | 19.5 | 4.3 | 0.0 | 9.1 |
| 1998 | 46.5 | 355.2 | 163.9 | 102.9 | 38.3 | 25.6 | 10.4 | 4.0 | 3.0 |
| 1999 | 143.2 | 311.7 | 201.0 | 64.8 | 37.4 | 18.1 | 11.1 | 9.4 | 12.1 |
| 2000 | 73.6 | 161.0 | 189.7 | 63.6 | 35.5 | 6.6 | 4.9 | 3.5 | 3.2 |
| 2001 | 55.3 | 164.8 | 145.6 | 47.1 | 5.9 | 21.5 | 2.3 | 7.4 | 0.0 |
| 2002 | 53.7 | 154.8 | 171.5 | 54.5 | 42.1 | 44.0 | 12.4 | 3.6 | 2.1 |
| 2003 | 73.7 | 165.8 | 65.3 | 29.1 | 5.9 | 14.8 | 10.4 | 1.5 | 3.7 |
| 2004 | 30.9 | 120.8 | 91.2 | 26.5 | 11.9 | 1.7 | 2.4 | 3.9 | 1.5 |
| 2005 | 25.2 | 70.9 | 77.4 | 47.7 | 22.4 | 12.6 | 3.7 | 0.0 | 1.2 |
| 2006 | 16.7 | 40.7 | 52.6 | 38.2 | 12.4 | 6.5 | 1.1 | 1.1 | 2.4 |
| 2007 | 47.0 | 136.0 | 60.7 | 22.2 | 17.1 | 4.1 | 2.2 | 0.4 | 0.7 |
| 2008 | 54.6 | 105.9 | 70.0 | 20.5 | 4.8 | 1.9 | 1.3 | 0.1 | 0.2 |
| 2009 | 13.4 | 112.0 | 78.4 | 30.3 | 10.7 | 4.7 | 0.0 | 0.8 | 0.6 |
| 2010 | 55.9 | 42.2 | 59.9 | 43.1 | 18.2 | 4.3 | 1.5 | 1.5 | 1.1 |
| 2011 | 18.7 | 83.2 | 53.9 | 35.5 | 22.2 | 10.7 | 3.7 | 0.8 | 1.3 |
| 2012 | 12.6 | 129.5 | 104.4 | 37.7 | 29.8 | 12.7 | 6.9 | 1.9 | 2.9 |
| 2013 | 4.6 | 35.3 | 67.6 | 25.5 | 6.2 | 3.9 | 2.4 | 0.9 | 0.4 |
| 2014 | 8.6 | 42.9 | 78.5 | 63.1 | 22.0 | 4.1 | 3.3 | 2.0 | 0.6 |
| 2015 | 4.6 | 11.3 | 15.9 | 13.3 | 9.6 | 3.5 | 0.6 | 0.7 | 0.3 |
| 2016 | 1.3 | 22.7 | 16.5 | 10.9 | 7.9 | 4.9 | 2.1 | 0.3 | 0.6 |
| 2017 | 8.1 | 8.5 | 37.8 | 20.2 | 13.0 | 10.0 | 7.0 | 2.5 | 1.8 |
| 2018 | 2.9 | 21.0 | 10.7 | 21.5 | 9.5 | 5.8 | 3.6 | 2.2 | 1.1 |

Table 27.4. Weight-at-age for plaice in 7.jk, 1993-2016.

|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0.196 | 0.256 | 0.306 | 0.417 | 0.582 | 0.751 | 0.939 | 1.151 | 1.707 |
| 1994 | 0.222 | 0.302 | 0.368 | 0.460 | 0.563 | 0.708 | 0.873 | 1.029 | 1.347 |
| 1995 | 0.228 | 0.272 | 0.325 | 0.391 | 0.521 | 0.651 | 0.840 | 0.817 | 1.546 |
| 1996 | 0.298 | 0.379 | 0.432 | 0.463 | 0.512 | 0.529 | 0.493 | 0.398 | 2.324 |
| 1997 | 0.295 | 0.339 | 0.430 | 0.483 | 0.654 | 0.807 | 0.937 | 0.669 | 1.319 |
| 1998 | 0.249 | 0.308 | 0.419 | 0.529 | 0.690 | 0.779 | 0.757 | 0.941 | 1.287 |
| 1999 | 0.289 | 0.354 | 0.417 | 0.596 | 0.627 | 0.840 | 0.882 | 1.170 | 1.382 |
| 2000 | 0.273 | 0.348 | 0.420 | 0.486 | 0.609 | 0.807 | 1.107 | 1.439 | 1.424 |
| 2001 | 0.243 | 0.325 | 0.405 | 0.537 | 0.644 | 0.800 | 0.550 | 1.115 | 0.000 |
| 2002 | 0.211 | 0.296 | 0.328 | 0.415 | 0.498 | 0.567 | 0.701 | 1.014 | 1.204 |
| 2003 | 0.274 | 0.358 | 0.402 | 0.482 | 0.575 | 0.734 | 0.876 | 1.041 | 1.646 |
| 2004 | 0.259 | 0.310 | 0.341 | 0.448 | 0.550 | 0.631 | 0.637 | 0.900 | 1.333 |
| 2005 | 0.238 | 0.276 | 0.324 | 0.381 | 0.459 | 0.731 | 0.949 | 0.845 | 1.615 |
| 2006 | 0.272 | 0.319 | 0.370 | 0.438 | 0.519 | 0.794 | 0.895 | 0.791 | 1.612 |
| 2007 | 0.239 | 0.281 | 0.354 | 0.433 | 0.482 | 0.573 | 0.727 | 1.394 | 1.108 |
| 2008 | 0.239 | 0.282 | 0.336 | 0.358 | 0.529 | 0.754 | 0.399 | 1.100 | 1.507 |
| 2009 | 0.224 | 0.255 | 0.335 | 0.403 | 0.462 | 0.520 | 0.569 | 1.080 | 1.266 |
| 2010 | 0.257 | 0.310 | 0.342 | 0.369 | 0.462 | 0.563 | 0.739 | 0.735 | 0.893 |
| 2011 | 0.257 | 0.282 | 0.321 | 0.355 | 0.407 | 0.626 | 0.625 | 0.507 | 0.984 |
| 2012 | 0.244 | 0.284 | 0.312 | 0.364 | 0.429 | 0.465 | 0.562 | 0.701 | 1.039 |
| 2013 | 0.256 | 0.294 | 0.336 | 0.400 | 0.462 | 0.503 | 0.609 | 0.744 | 1.002 |
| 2014 | 0.250 | 0.288 | 0.321 | 0.377 | 0.425 | 0.471 | 0.526 | 0.609 | 0.992 |
| 2015 | 0.295 | 0.349 | 0.378 | 0.439 | 0.509 | 0.565 | 0.645 | 0.611 | 0.743 |
| 2016 | 0.344 | 0.364 | 0.433 | 0.484 | 0.528 | 0.584 | 0.677 | 0.686 | 0.737 |
| 2017 | 0.320 | 0.357 | 0.423 | 0.491 | 0.570 | 0.607 | 0.685 | 0.713 | 0.787 |
| 2018 | 0.324 | 0.385 | 0.425 | 0.501 | 0.558 | 0.624 | 0.748 | 0.760 | 0.832 |

Table 27.5. Tuning data. The ages and years used in the assessment are in bold.


## Table 27.6. XSA diagnostics.

```
FLR XSA Diagnostics 2019-05-26 08:45:33
CPUE data from indices
```

Catch data for 26 years 1993 to 2018. Ages 4 to 8 .
fleet first age last age first year last year alpha beta
1 IRL-VMS: nos per 1000 hours $4 \quad 7 \quad 20062018$ <NA> <NA>
Time series weights :
Tapered time weighting not applied
Catchability analysis :
Catchability independent of size for all ages
Catchability independent of age for ages > 6
Terminal population estimation :
Survivor estimates shrunk towards the mean $F$
of the final 5 years or the 3 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1$
Minimum standard error for population
estimates derived from each fleet $=0.3$
prior weighting not applied
Regression weights
year
age 2009201020112012201320142015201620172018
$\begin{array}{lllllllllll}\text { al1 } & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$
Fishing mortalities
year
age $2009 \quad 2010 \quad 2011 \quad 2012 \quad 2013 \quad 2014 \quad 2015 \quad 2016 \quad 2017 \quad 2018$

```
    4 0.560 0.512 0.635 0.936 0.524 0.963 0.313 0.316 0.561 0.562
    5 0.504 0.626 0.593 1.201 0.556 1.290 0.370 0.334 0.721 0.658
    6 0.801 0.587 0.704 1.450 0.564 1.285 0.601 0.357 0.761 0.821
    7 0.860 0.812 0.753 1.078 0.657 0.832 0.633 0.642 0.947 0.852
    8 0.860 0.812 0.753 1.078 0.657 0.832 0.633 0.642 0.947 0.852
```

XSA population number (Thousand)
age
$\begin{array}{llllll}\text { year } & 4 & 5 & 6 & 7 & 8\end{array}$
$20091948121 \quad 9 \quad 3$
$\begin{array}{lllll}2010 & 159 & 98 & 44 & 8 \\ 8\end{array}$
201112284472112
201218257412019
$\begin{array}{llllll}2013 & 176 & 63 & 15 & 9 & 8\end{array}$
20141359232811
$2015 \quad 63 \quad 46 \quad 23 \quad 8 \quad 4$
$2016 \quad 6541 \quad 2811 \quad 7$
$2017 \quad 9442 \quad 261719$
$2018 \quad 2647181113$
Estimated population abundance at 1st Jan 2019
age
$\begin{array}{llllll}\text { year } & 4 \quad 5 \quad 6 & 7 & 8\end{array}$
20190132274
Fleet: IRL-VMS: nos per 1000 hours
Log catchability residuals.


# Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time 


, Age 5 Year class =2013
source

|  | scaledwts survivors yrcls |  |  |
| :--- | ---: | ---: | ---: |
| IRL-VMS: nos per 1000 hours | 0.747 | 22 | 2013 |
| fshk | 0.253 | 22 | 2013 |

, Age 6 Year class $=2012$
source

|  | scaledwts | survivors | yrcls |
| :--- | ---: | ---: | ---: |
| IRL-VMS: nos per 1000 hours | 0.731 | 8 | 2012 |
| fshk | 0.269 | 9 | 2012 |

, Age 7 Year class =2011
source

|  | scaledwts survivors | yrcls |  |
| :--- | ---: | ---: | ---: |
| IRL-VMS: nos per 1000 hours | 0.826 | 5 | 2011 |
| fshk | 0.174 | 6 | 2011 |

Table 27.7. Summary table for ple-7.jk assessment in absolute values. Landings in tonnes. Recruitment (age 4) in thousands. SSB in tonnes.

| Year | Land 7h-k | Land 7.jk | Recruit | Fbar | SSB |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 672 | 437 | 726 | 0.932 | 400 |
| 1994 | 578 | 317 | 508 | 0.743 | 355 |
| 1995 | 667 | 419 | 650 | 0.724 | 362 |
| 1996 | 560 | 336 | 485 | 0.708 | 374 |
| 1997 | 551 | 375 | 480 | 0.788 | 409 |
| 1998 | 454 | 306 | 378 | 0.777 | 352 |
| 1999 | 385 | 311 | 392 | 0.822 | 383 |
| 2000 | 337 | 225 | 361 | 0.616 | 314 |
| 2001 | 313 | 182 | 234 | 0.533 | 277 |
| 2002 | 316 | 183 | 265 | 1.234 | 203 |
| 2003 | 237 | 147 | 151 | 0.667 | 154 |
| 2004 | 232 | 104 | 181 | 0.555 | 126 |
| 2005 | 170 | 94 | 158 | 0.927 | 116 |
| 2006 | 143 | 71 | 101 | 0.894 | 96 |
| 2007 | 142 | 94 | 115 | 1.059 | 73 |
| 2008 | 122 | 79 | 166 | 0.558 | 78 |
| 2009 | 148 | 80 | 194 | 0.622 | 106 |
| 2010 | 156 | 78 | 159 | 0.575 | 114 |
| 2011 | 179 | 80 | 122 | 0.644 | 104 |
| 2012 | 199 | 112 | 182 | 1.196 | 110 |
| 2013 | 141 | 65 | 176 | 0.548 | 93 |
| 2014 | 153 | 89 | 135 | 1.179 | 96 |
| 2015 | 108 | 33 | 63 | 0.428 | 59 |
| 2016 | 100 | 37 | 65 | 0.336 | 70 |
| 2017 | 116 | 53 | 94 | 0.681 | 94 |
| 2018 | 95 | 38 | 26 | 0.68 | 60 |
| 2019 |  |  | 216 |  | 109 |



Figure 27.1.a. The spatial distribution of International landings of Plaice (2012 data, all gears combined; data from STECF).




Figure 27.1b. Landings, Lpue and effort for Irish otter trawlers in 7.j.
ple.27.7jk


Age

Figure 27.3. Age distribution of plaice landings in 7.jk between 1993 and 2018. All gears and quarters combined. The age data for 1995 were considered insufficient and for this year the combined age data for 19931996 were used.
ple.27.7jk
Standardised landings proportions-at-age


Figure 27.4. Standardised landings proportions-at-age for plaice in 7.jk. Grey bubbles represent higher than average catch-at-age and black bubbles represent lower than average catch-at-age.

Plaice 27.7j-k
stock weights


Figure 27.5. Landings weights / stock weights of ple7jk.


Figure 27.6. Top: the proportion of plaice in landings of Irish vessels with VMS over the years 2006-2016. The black line indicates the polygon inside which plaice are caught. Effort and landings from the VMS/logbooks data inside the polygon were used as a tuning index. Bottom: the VMS lpue index (black line) and the lpue of plaice in the whole of 7.j.



Figure 27.7. The log-standardised tuning index by year (top) and cohort (bottom). Due to the lack of contrast in the numbers-at-age cohorts are not tracked particularly well.

IRL-VMS: nos per 1000 hours

log index

Figure 27.8. Internal consistency of the tuning fleet.

## Residuals <br> Plaice 27.7j-k

IRL-VMS: nos per 1000 hours


Figure 27.9. Residuals of the index fit.


Figure 27.10. Retrospective analysis of the assessment.


Figure 27.11. Ple7.jk stock-recruit plot, estimated as part of WGCSE 2017. Because recruitment does not appear to be impaired at the lowest stock size, the inflection point of the segmented regression was chosen to be the lowest biomass that generated high recruitment.


Figure 27.12. Ple7.jk Summary of MSY evaluations (without $B_{\text {trigger }}$ harvest control rule), a) simulated and observed recruitment, b)simulated and observed biomass, c) simulated an observed catch and d) Cumulative probability of $F_{\text {msy }}$ and $S S B<B_{\text {lim }}$ and $B_{\text {pa., }}$, all estimated as part of WGCSE 2017.


Figure 27.13. Ple7.jk Summary of MSY evaluations (with $B_{\text {trigger }}=B_{\text {lim }}$ harvest control rule), a) simulated and observed recruitment, b)simulated and observed biomass, c) simulated an observed catch and d) Cumulative probability of $F_{\text {msy }}$ and $S S B<B_{l i m}$ and $B_{\text {pa., }}$, estimated as part of WGCSE 2017.

## 27 Pollack (Pollachius pollachius) in subareas 6-7 (Celtic Seas and the English Channel)

## Type of assessment in 2019

The Celtic Sea and West of Scotland (Subareas6 and7) Pollackstock is considered a Data Limited Stock, classified by ICES WKLIFE II (ICES CM2012/ACOM:79) as category 4.1.2. DCAC (Deple-tion-Corrected Average Catch) method is recommended to assess this stock, which is performed through the NOAA toolbox.

## ICES advice applicable to 2020

ICES advices that when the precautionary approach is applied, commercial catches should not exceed 3360 tonnes in 2020.

### 27.1 General

### 27.1.1 Stock Identity

This section is not dedicated to a 'stock', it relates to a species in a wider region where data are available. The stock structure of Pollack populations in this ecoregion is not clear. ICES does not necessarily advocate that subareas 6 and 7 constitutes a management unit for Pollack, and further work is required.

### 27.1.2 Management applicable to 2019

The 2019 TAC for Pollack is set for ICES subareas 6 (and 5a, b; international waters of 12 and 14) and 7 separately.

| Species: | Pollack <br> Pollachius pollachius |  | Zone: | 6; Union and international waters of 5 b; international waters of 12 and 14 (POL/56-14) |
| :---: | :---: | :---: | :---: | :---: |
| Spain |  | 6 |  |  |
| France |  | 190 |  |  |
| Ireland |  | 56 |  |  |
| United Kingdom |  | 145 |  |  |
| Union |  | 397 |  |  |
| TAC |  | 397 |  | Precautionary TAC |
| Species: | Pollack <br> Pollachius pollachius |  | Zone: | $\begin{aligned} & 7 \\ & \text { (POL/07.) } \end{aligned}$ |
| Belgium |  | $378{ }^{(2)}$ |  |  |
| Spain |  | 23 (2) |  |  |
| France |  | $8712{ }^{(1)}$ |  |  |
| Ireland |  | 929 ( ${ }^{(1)}$ |  |  |
| United Kingdom |  | $2121{ }^{(1)}$ |  |  |
| Union |  | 12163 (2) |  |  |
| TAC |  | 12163 |  | Precautionary TAC <br> Article 13(1) of this Regulation applies |

${ }^{\left({ }^{()}\right)}$Special condition: of which up to $2 \%$ may be fished in: $8 \mathrm{Ba}, 8 \mathrm{~b}, \mathrm{8d}$ and Se (POL/ $\left./ 8 \mathrm{SABE}\right)$.

The 2018 TAC uptake for Subarea 6 was low at $16 \%$ and varied considerably between countries. France, which holds $48 \%$ of the TAC, only utilized $0.03 \%$ of their quota. The UK utilized $19.7 \%$ of the $37 \%$ TAC allocation, Ireland had the largest quota uptake at $60.5 \%$ constituting $14 \%$ of the TAC allocation and finally Spain utilized $6.8 \%$ of the $1 \%$ TAC allocation.
In Subarea7, the uptake was slightly higher at $23 \%$ and again varied considerably between countries. France which holds the majority of the TAC allocation (71.6\%), only utilized 9\% of this. The UK utilized $59 \%$ of its $17.4 \%$ TAC allocation, Ireland utilized $80 \%$ of its $7.6 \%$ TAC allocation, Belgium and Spain which hold very low TAC allocations at $3.1 \%$ and $0.2 \%$, utilized $5 \%$ and $54 \%$ respectively.

## Fishery in 2018

## Landings

2891 tonnes of pollack were landed in $2018,98 \%$ of which came from subarea 7 . The nominal landings for ICES subareas 6 and 7 are shown in Tables 28.1 and 28.2 respectively. For Subarea 6 , there was a $30.2 \%$ increase in landings ( 63 tonnes) in 2018 compared to the landings in 2017 (44 tonnes). Ireland declared the highest landings (54\%) followed by the UK (46\%), Spain ( $0.6 \%$ ) and France ( $0.1 \%$ ) respectively There was a $13.8 \%$ decrease in landings ( 2828 tonnes) for Subarea 7 in 2018 compared to 2017 ( 3260 tonnes). The UK had the highest landings ( $44 \%$ ) followed by France (29\%), Ireland (26\%), Belgium (0.7\%) and Spain (0.4\%).

## Landings by division

In Subarea 6, all landings in 2018 derived from Division 6.a (2\% of the overall breakdown). In Subarea 7, the division with the highest proportion of landings derived from 7.e (40\%) followed by 7.j (27\%), 7.h (13\%) and 7.f (9\%). Landings in divisions 7.a, b, c, d, g and h were negligible (8.9\%).

## Landings by gear

The majority of Pollack landings in the Celtic Sea ecoregions are caught by gillnets (43.9\%) followed by bottom trawlers (23.4\%) trolling lines (16.3\%), miscellaneous gears (13.1\%), beam trawlers ( $2.6 \%$ ) and seiners ( $0.7 \%$ ). When separated by subarea, the predominant gears landing pollack in Subarea 6 are bottom trawlers (59.2\%) followed by gillnets ( $32.7 \%$ ), miscellaneous gears ( $7.1 \%$ ) and trolling lines (1\%). In Subarea 7, gillnets have the highest landings (44.1\%) followed by bottom trawlers ( $22.6 \%$ ) trolling lines (16.7\%), miscellaneous gears (13.2\%), beam trawlers (2.6\%) and seiners (0.7\%).

## Landings by quarter

Pollack are not targeted throughout the entire year, and are mainly targeted during the first quarter which coincides with spawning. The breakdown of landings per quarter shows that the highest landings were in quarter $1(40 \%)$ followed by quarter $2(27 \%)$, quarter $3(16 \%)$ and quarter $4(17 \%)$ respectively. As France, UK and Ireland constitute $99 \%$ of the pollack landings, the landings per quarter 1 were ( $36 \%, 47 \%, 32 \%$ ), quarter 2 ( $33 \%, 22 \%, 30 \%$ ), quarter $3(19 \%, 13 \%, 19 \%)$ and quarter 4 ( $12 \%, 19 \%, 19 \%$ ) respectively.

## Discards

Discarding is negligible at 19.9 tonnes. $91.9 \%$ of which coming from gillnetters, followed by bottom trawls at $7.9 \%$ and beam trawls at $0.2 \%$.

## Landings uncertainty

Pollack is a known recreational fishing species, however; it is unknown as to the quantities exploited by recreational fisheries. A phone study conducted in France in 2011-2013 by Levrel et al. (2013) estimated that 3300 tonnes are landed annually through recreational fishing, 2274 tonnes of which are retained. Radford et al., 2018 further suggest that pollack landings may be similar to or above commercial landings.

### 27.2 Stock assessment

A DCAC (Depleted-Corrected Adjusted Catch) method is used to estimate a yield likely to be sustainable (MacCall, 2009). TheDCAC-method was applied during WGCSE 2019 with the same model settings as applied the previous year's assessment (ICES, 2018).

Subareas 6 and 7 are run independently. For Subarea 6; six separate model runs using various parameters are conducted giving an average DCAC value plus an upper and lower $95 \%$ confidence interval and for Subarea 7; nine separate model runs using various parameters are conducted given an average DCAC value plus an upper and lower $95 \%$ confidence interval.

The information provided for the assessment is insufficient to evaluate the exploitation and the trends of pollack in the Celtic Seas ecoregion. Commercial catches have declined since the late 1980s, and in 2018 are the lowest in the time-series.

The input data and parameters used for the assessment are detailed in Tables 28.3,28.4.

## 2019 Results

The average DCAC values (Figure 28.1) show that in both subareas 6 and 7, landings are below the average DCAC by 85 tonnes in Subarea 6 and 1182 tonnes in Subarea 7. This suggests that yield in Subarea 6 could be increased up to 148 tonnes and 4010 tonnes in Subarea 7.

## Comparison with previous assessment

Table 28.5 compares the results with the previous year's assessment. The results are consistent with the range of DCAC values estimated when the method was previously applied.

## Uncertainties in assessment and forecast

The DCAC model relies solely on commercial catch data and does not include any biological or survey data that are available for this stock. The model also cannot estimate reference points.
By construction, the DCAC method only uses long time-series of official landings. As the output is a smoothed value of the landings over the assessed time series, the computations of DCAC are always similar to the previous year's results, even when recruitment or SSB fluctuate. To test the model, during WGCSE 2019, proxy values were used to demonstrate the issues with the model results computed annually. Two separate tests were conducted, the first to substantially increase the landings and the second to crash the stock. In the first test run, 1000 tonnes were added to Subarea 6 timeline which gave an average DCAC of 169 tonnes (increase of 21 tonnes from the actual 2018 of 148 tonnes) and in Subarea7,12 000 tonnes were added to the timeline which gave an average DCAC of 4228 tonnes (increase of 218 tonnes from the actual 2018 of 4010 tonnes). The second test run added 0 tonnes to Subarea 6 timeline which gave an average DCAC of 146 tonnes (decrease of 2 tonnes from the actual 2018 of 148 tonnes) and in Subarea 7, 0 tonnes were added to the timeline which gave an average DCAC of 3972 tonnes (decrease of 38 tonnes from the actual 2018 of 4010 tonnes) (Table 28.6). This highlights that the DCAC model will not take any account of the state of the stock.

Recreational catch is unknown and therefore cannot be estimated or incorporated into the assessment. From preliminary data, it seems likely that catches in recreational fisheries are of a similar order of magnitude to, or larger than, commercial landings.

### 27.3 Management considerations

TAC for Subarea 7 includes ICES Division 7d, which is not in the remit of the Celtic Sea ecoregion. TAC set for both subarea 6 and 7 are not in line with the current estimates of catches and estimated sustainable yields, and therefore are not constraining.

### 27.3.1 Management plan

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to demersal stocks including pollack in ICES subareas 6 and 7.

### 27.3.2 Recommendations

DCAC continues to be the reference model for the Pollack assessment in the coming year. However; WGCSE considers that pol. 27.67 should be assessed using a different model therefore different assessment models shall be explored over the coming year with a purpose to Benchmark in 2021.

### 27.4 References

ICES. 2012. Report of The Workshop to Finalize the ICES Data-limited Stock (DLS) Methodologies Documentation in an Operational Form for the 2013 Advice Season and to make Recommendations on Target Categories for Data-limited Stocks (WKLIFE II), 20-22 November 2012, Copenhagen, Denmark. ICES CM 2012/ACOM:79. 46 pp.
ICES. 2018. Advice basis. In Report of the ICES Advisory Committee, 2018. ICES Advice 2018, Book 1, Section 1.2. https://doi.org/10.17895/ices.pub. 4503.

Levrel H., Bellanger M., Le Goff R. and Drogou M. 2013. La pêche récréative en mer en France métropolitaine (Atlantique, Manche, Mer du Nord, Méditerranée). Résultats de l'enquête 2011-2013. http://archimer.ifremer.fr/doc/00162/27300/.

MacCall, A. D. 2009. Depletion-corrected average catch: a simple formula for estimating sustainable yields in data-poor situations. - ICES Journal of Marine Science, 66: 2267-2271.

Radford Z, Hyder K, Zarauz L, Mugerza E, Ferter K, Prellezo R, et al. 2018. The impact of marine recreational fishing on key fish stocks in European waters. PLoS ONE 13(9): e0201666. https://doi.org/10.1371/journal.pone. 0201666 .

Table 28.1. Landings of Pollack in subarea 6 as officially reported to ICES.

| Year | Belgium | Denmark | France | Germany | Ire- <br> land | Netherlands | Norway | Portugal | Spain | Sweden | UK | Total <br> Subarea 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 1 | - | - | - | - | - | - | - | - | - | 295 | 296 |
| 1951 | - | - | - | - | - | - | - | - | - | - | 484 | 484 |
| 1952 | - | - | - | - | - | 1 | - | - | - | - | 503 | 504 |
| 1953 | - | - | - | - | - | - | - | - | - | - | 422 | 422 |
| 1954 | - | - | - | - | - | - | - | - | - | - | 452 | 452 |
| 1955 | - | - | - | - | - | - | - | - | - | - | 566 | 566 |
| 1956 | - | - | - | - | - | - | - | - | - | - | 528 | 528 |
| 1957 | - | - | - | - | - | - | - | - | - | - | 547 | 547 |
| 1958 | . | - | - | 23 | - | - | - | - | - | - | 710 | 733 |
| 1959 | 1 | - | - | 6 | - | - | - | - | - | - | 607 | 614 |
| 1960 | 15 | - | - | - | - | - | - | - | - | - | 441 | 456 |
| 1961 | 1 | - | - | 1 | 125 | - | - | - | - | - | 259 | 386 |
| 1962 | 2 | - | - | 8 | 197 | - | - | - | - | - | 235 | 442 |
| 1963 | 6 | - | - | 2 | 204 | - | - | - | - | - | 320 | 532 |
| 1964 | 1 | - | - | 1 | 130 | - | - | - | - | - | 368 | 500 |
| 1965 | 1 | - | - | 1 | 402 | - | - | - | - | - | 496 | 900 |
| 1966 | 2 | - | - | - | 200 | - | - | - | - | - | 428 | 630 |
| 1967 | 1 | - | - | 1 | 263 | - | - | - | - | 1106 | 413 | 1784 |
| 1968 | 5 | - | - | 2 | 214 | - | 148 | - | - | 1012 | 500 | 1881 |
| 1969 | 1 | - | - | 4 | 282 | - | - | - | - | 1224 | 667 | 2178 |
| 1970 | 2 | - | - | 1 | 398 | - | - | - | - | 756 | 447 | 1604 |
| 1971 | 1 | - | - | 5 | 75 | - | - | - | - | 750 | 256 | 1087 |
| 1972 | 1 | - | - | 1 | 127 | - | - | - | - | 779 | 317 | 1225 |
| 1973 | 2 | - | - | - | - | - | - | - | - | - | 503 | 505 |
| 1974 | 6 | - | - | - | - | 3 | - | - | - | - | 359 | 368 |
| 1975 | < 0.5 | - | - | 1 | - | 1 | 4 | - | - | - | 393 | 399 |
| 1976 | 7 | - | - | - | - | 1 | - | - | - | - | 519 | 527 |
| 1977 | - | - | 196 | - | - | 1 | 2 | - | - | - | 493 | 692 |
| 1978 | - | - | 196 | - | - | - | 4 | - | - | - | 553 | 753 |
| 1979 | - | - | 310 | - | - | - | - | - | - | - | 350 | 660 |
| 1980 | - | - | 36 | - | - | - | - | - | - | - | 233 | 269 |
| 1981 | - | - | 342 | - | - | - | - | - | 55 | - | 185 | 582 |
| 1982 | - | < 0.5 | 272 | - | - | - | - | - | 95 | - | 103 | 470 |
| 1983 | - | - | 331 | - | - | - | - | - | 86 | - | 148 | 565 |
| 1984 | - | - | 212 | - | - | - | - | - | 222 | - | 194 | 628 |
| 1985 | $<0.5$ | - | 224 | 1 | - | - | - | - | 283 | - | 328 | 836 |
| 1986 | - | - | 145 | - | 223 | - | - | - | 2217 | - | 187 | 2772 |


| Year | Belgium | Denmark | France | Germany | Ire- land | Netherlands | Norway | Portugal | Spain | Sweden | UK | Total <br> Subarea 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | - | < 0.5 | 108 | - | 103 | - | - | - | 860 | - | 259 | 1330 |
| 1988 | - | < 0.5 | 128 | - | 163 | - | - | - | 1925 | - | 221 | 2437 |
| 1989 | - | < 0.5 | 111 | 1 | 103 | - | - | - | - | - | 179 | 394 |
| 1990 | - | - | 76 | - | 150 | - | 1 | - | - | - | 192 | 419 |
| 1991 | - | - | 31 | - | 145 | - | - | - | 4 | - | 189 | 369 |
| 1992 | - | <0.5 | 21 | - | 23 | - | - | - | <0.5 | - | 203 | 247 |
| 1993 | - | - | 39 | - | 12 | - | - | - | - | - | 273 | 324 |
| 1994 | - | - | 34 | $<0.5$ | 26 | - | < 0.5 | - | - | - | 276 | 336 |
| 1995 | - | - | 64 | 3 | 83 | $-$ | - | - | - | - | 354 | 504 |
| 1996 | - | <0.5 | 29 | <0.5 | 97 | - | 1 | - | - | - | 210 | 337 |
| 1997 | - | - | 14 | 1 | 69 | - | 2 | - | - | - | 162 | 248 |
| 1998 | - | - | 21 | - | 60 | - | - | < 0.5 | < 0.5 | - | 147 | 228 |
| 1999 | - | - | - | - | 73 | - | 3 | - | <0.5 | - | 136 | 212 |
| 2000 | - | - | 11 | 2 | 62 | - | - | - | - | - | 116 | 191 |
| 2001 | - | - | 8 | - | 108 | - | - | - | - | - | 101 | 217 |
| 2002 | - | - | 9 | - | 26 | - | - | - | - | - | 96 | 131 |
| 2003 | < 0.5 | - | 3 | - | 88 | - | 1 | - | - | - | 111 | 203 |
| 2004 | <0.5 | - | 2 | - | 68 | - | 1 | - | - | - | 65 | 136 |
| 2005 | - | - | 23 | - | 28 | - | - | - | - | - | 16 | 67 |
| 2006 | - | - | 3 | - | 25 | - | < 0.5 | - | 4 | - | 5 | 37 |
| 2007 | - | - | 10 | - | 21 | - | 6 | - | - | - | 21 | 58 |
| 2008 | $-$ | - | 8 | - | 21 | - | 1 | - | - | - | 23 | 53 |
| 2009 | - | - | 7 | - | 5 | - | <0.5 | - | - | - | 25 | 37 |
| 2010 | - | - | 6 | - | 34 | - | < 0.5 | - | - | - | 38 | 78 |
| 2011 | - | - | 3 | - | 8 | - | - | - | - | - | 34 | 45 |
| 2012 | - | - | 2 | - | 10 | - | - | - | - | - | 33 | 45 |
| 2013 | - | - | 1 | - | 34 | - | - | - | - | - | 22 | 57 |
| 2014 | - | - | 1 | - | 25 | - | - | - | - | - | 18 | 44 |
| 2015 | - | - | $<0.5$ | - | 23 | - | $<0.5$ | - | - | - | 25 | 48 |
| 2016 | - | $-$ | <0.5 | - | 44 | - | <0.5 | - | - | - | 29 | 74 |
| 2017* | $-$ | - | $<0.5$ | - | 30 | - | < 0.5 | - | - | - | 14 | 44 |
| 2018* | $-$ | - | <0.5 | - | 34 | $-$ | <0.5 | - | $<0.5$ | - | 28 | 63 |

Table 28.2. Landings of Pollack in subarea 7 as officially reported to ICES.

| Year | Belgium | Denmark | France | Germany | Ireland | Netherlands | Norway | Spain | UK | Total |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |  |
| 19 | - | - | - | - | - | - | - | - | - | - | - |


| Year | Belgium | Denmark | France | Germany | Ireland | Netherlands | Norway | Spain | UK | Total <br> Subarea <br> 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 241 | - | 4574 | - | 1335 | - | - | 17 | 1795 | 7962 |
| 1987 | 149 | - | 5213 | - | 848 | - | - | 19 | 2010 | 8239 |
| 1988 | 191 | - | 5211 | - | 1066 | - | - | 22 | 1740 | 8230 |
| 1989 | 145 | - | 3893 | - | 994 | - | - | 18 | 1487 | 6537 |
| 1990 | 133 | - | 4831 | - | 1066 | - | - | 26 | 1914 | 7970 |
| 1991 | 76 | - | 3211 | - | 1045 | - | - | 22 | 1962 | 6316 |
| 1992 | 62 | - | 2849 | - | 1014 | - | - | 19 | 1889 | 5833 |
| 1993 | 55 | - | 2325 | - | 1137 | - | - | 7 | 2135 | 5659 |
| 1994 | 94 | - | 2621 | - | 921 | - | - | 8 | 2391 | 6035 |
| 1995 | 88 | 2 | 2315 | - | 1107 | - | - | 4 | 2168 | 5684 |
| 1996 | 94 | - | 2684 | - | 1190 | 6 | - | 5 | 2519 | 6498 |
| 1997 | 99 | - | 2443 | - | 984 | 4 | < 0.5 | 7 | 2540 | 6077 |
| 1998 | 92 | - | 2375 | - | 886 | 1 | - | 11 | 2347 | 5712 |
| 1999 | 86 | - | - | - | 976 | - | 3 | 19 | 1703 | 2787 |
| 2000 | 71 | - | 2422 | - | 1069 | - | - | 5 | 1810 | 5377 |
| 2001 | 100 | - | 2515 | - | 1274 | - | - | 9 | 1987 | 5885 |
| 2002 | 117 | - | 2481 | - | 1308 | - | - | 17 | 1999 | 5922 |
| 2003 | 113 | - | 2284 | - | 1151 | - | - | 12 | 1788 | 5348 |
| 2004 | 104 | - | 1914 | - | 1049 | 1 | - | 13 | 1705 | 4786 |
| 2005 | 98 | - | 2198 | - | 728 | 1 | - | 16 | 1684 | 4725 |
| 2006 | 79 | - | 2213 | - | 809 | 1 | - | 28 | 1513 | 4643 |
| 2007 | 91 | - | 1970 | - | 782 | 3 | - | 1 | 1764 | 4611 |
| 2008 | 76 | - | 1579 | - | 738 | 1 | - | 14 | 1453 | 3861 |
| 2009 | 42 | - | 1670 | - | 828 | 4 | - | 3 | 1545 | 4092 |
| 2010 | 35 | - | 1846 | - | 942 | 2 | - | 3 | 1459 | 4284 |
| 2011 | 28 | - | 1415 | - | 912 | 1 | - | - | 1716 | 4072 |
| 2012 | 43 | - | 1421 | - | 1165 | 1 | - | 3 | 1835 | 4468 |
| 2013 | 39 | - | 1790 | - | 1249 | 1 | - | 11 | 1838 | 4928 |
| 2014 | 84 | - | 2042 | - | 1096 | 1 | - | 14 | 2122 | 5359 |
| 2015 | 32 | - | 1154 | - | 1070 | 1 | - | 13 | 1469 | 3739 |
| 2016 | 42 | - | 1237 | - | 1073 | < 0.5 | - | 12 | 1842 | 4206 |
| 2017* | 19 | - | 959 | - | 891 | < 0.5 | - | 14 | 1377 | 3260 |
| 2018* | 21 | - | 823 | - | 741 | - | - | 12 | 1231 | 2828 |

Table 28.3. Input parameters for the six DCAC runs carried out for Pollack in Subarea 6.

|  | Run 1 | Run 2 | Run 3 | Run 4 | Run 5 | Run 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| sumC | 6782 | 6782 | 6782 | 6782 | 6782 | 6782 |
| CV sumC | 0 | 0 | 0 | 0 | 0 | 0 |
| no of years | 33 | 33 | 33 | 33 | 33 | 33 |
| iterations | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| stdev M | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 |
| $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ | 0.6 | 0.8 | 1 | 0.6 | 0.8 | 1 |
| stdev $\mathrm{F}_{\text {MSY }}$ to M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| distr $\mathrm{F}_{\text {MSY }}$ to M | normal | normal | normal | normal | normal | normal |
| $\mathrm{B}_{\mathrm{MSY}} / \mathrm{BO}$ | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 |
| stdev $\mathrm{B}_{\text {Mš }} / \mathrm{BO}$ | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| up $\lim \mathrm{B}_{\text {MSV }} / \mathrm{BO}$ | 1 | 1 | 1 | 1 | 1 | 1 |
| low lim $\mathrm{B}_{\mathrm{Msr}} / \mathrm{BO}$ | 0 | 0 | 0 | 0 | 0 | 0 |
| depletion delta $\Delta$ | 0.8 | 0.8 | 0.8 | 0.9 | 0.9 | 0.9 |
| stdev $\Delta$ | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| distr $\Delta$ | normal | normal | normal | normal | normal | normal |

Table 28.4. Input parameters for the 9 DCAC runs carried out for Pollack in Subarea 7.

|  | Run 1 | Run 2 | Run 3 | Run 4 | Run 5 | Run 6 | Run 7 | Run 8 | Run 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| sumC | 168434 | 168434 | 168434 | 168434 | 168434 | 168434 | 168434 | 168434 | 168434 |
| CV sumC | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| no of years | 33 | 33 | 33 | 33 | 33 | 33 | 33 | 33 | 33 |
| iterations | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| stdev M | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 |
| $\mathrm{F}_{\text {MSY }} / \mathrm{M}$ | 0.6 | 0.8 | 1 | 0.6 | 0.8 | 1 | 0.6 | 0.8 | 1 |
| stdev $\mathrm{F}_{\text {MSY }}$ to M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| distr $\mathrm{F}_{\text {MSY }}$ to M | normal | normal | normal | normal | normal | normal | normal | normal | normal |
| $\mathrm{B}_{\text {MSY }} / \mathrm{BO}$ | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 |
| stdev $\mathrm{B}_{\text {Ms/ }} / \mathrm{BO}$ | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| up $\lim \mathrm{B}_{\text {MSY }} / \mathrm{BO}$ | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| low lim $\mathrm{B}_{\text {MsY }} / \mathrm{BO}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| depletion delta $\Delta$ | 0.5 | 0.5 | 0.5 | 0.6 | 0.6 | 0.6 | 0.7 | 0.7 | 0.7 |
| stdev $\Delta$ | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| distr $\Delta$ | normal | normal | normal | normal | normal | normal | normal | normal | normal |

Table 28.5. Comparison of the 2019 DCAC assessment and previous DCAC results.

| Year | Subarea 6 landings (tonnes) | Average DCAC value | Subarea 7 landings (tonnes) | Average DCAC value |
| :--- | :---: | :---: | :---: | :---: | :---: |
| 2018 | 63 | 148 | 2891 | 4010 |
| 2017 | 44 | 150 | 3260 | 4042 |
| 2016 | 74 | 152 | 4131 | 4063 |
| 2015 | 48 | 155 | 5359 | 4062 |
| 2014 | 44 | 156 | 4468 | 3953 |
| 2013 | 57 |  |  | 4020 |

Table 28.6. Test DCAC runs compared to previous DCAC results.

|  | Subarea 6 landings <br> (tonnes) | Average DCAC value | Subarea 7 landings <br> (tonnes) | Average DCAC value |
| :--- | :---: | :---: | :---: | :---: |
| Test 1 | 1000 | 169 | 12000 | 4228 |
| Test 2 | 0 | 146 | 0 | 3972 |
| 2018 | 63 | 148 | 2891 | 4010 |
| 2017 | 44 | 150 | 3260 | 4042 |
| 2016 | 48 | 152 | 3740 | 4063 |
| 2015 | 44 | 156 | 5359 | 4020 |
| 2014 | 57 | 158 | 4468 | 3953 |
| 2013 |  |  |  | 4062 |



Figure 28.1. Pollack in subareas 6 and 7. The results of the depletion-corrected average catch (DCAC) assessment method as applied to commercial landings data since 1986. The grey box indicates the proxy for the maximum sustainable catch $\pm 95 \%$ confidence intervals.

## 28 Saithe (Pollachius virens) in subareas 7-10

## Type of assessment in 2019

Saithe was included as a stock for assessment in WGCSE 2019, with the view to provide advice for this stock for 2020. Saithe in subareas $7-10$ is currently considered as a data-limited stock as only landings data have been made available over time, and was classified as a category 4.1.2 stock during WGCSE 2019. A DCAC (Depletion-Corrected Average Catch) method was therefore recommended to assess this stock, which is performed through the NOAA toolbox.

## ICES advice applicable to 2020

ICES advises that when the precautionary approach is applied, commercial catches should be no more than 586 tonnes in each of the years 2020, 2021 and 2022.

The above advice was established by averaging the catches from 2016-2018 (732 tonnes). Further to this, a $20 \%$ precautionary buffer was applied as the stock size is unknown.

### 28.1 General

### 28.1.1 Stock Identity

The stock structure of saithe populations in subareas $7-10$ is unclear, and suggestions have been forwarded that saithe populations are inclusive of the entire Northeast Atlantic region. Figure 31.1 displays the landings from 27.3a.4.6 and 27.7-10. Further work and consideration is therefore required regarding the management of saithe.

### 28.1.2 Management applicable to 2019

The 2019 TAC for saithe is set for ICES subareas 7-10.

| Species:Saithe <br> Pollachius virens  Zone: <br> Belgium 6 $7,8,9$ and 10; Union waters of CECAF 34.1.1 <br> $($ POK $/ 7 / 3411)$ <br> France 1245  <br> Ireland 1491  <br> United Kingdom 434  <br> Union 3176  <br> TAC 3176 Precautionary TAC |
| :--- | ---: | :--- | :--- |

The 2018 TAC uptake in subareas 7-10 was low at $15.9 \%$ and varied considerably between countries. Ireland, which holds $47 \%$ of the TAC, utilised $24 \%$ of their quota. France utilised $5 \%$ of the $39 \%$ TAC allocation, the UK utilised $18 \%$ of their $14 \%$ TAC allocation and finally Belgium utilised $23 \%$ of the $0.002 \%$ TAC allocation.

## Fishery in 2018

The ICES preliminary data reported 496 tonnes of saithe were landed in 2018 in ICES subareas 7-10.

The data extracted from InterCatch only contained data from Ireland; 365 tonnes, UK England, Wales and Scotland; 54 tonnes and the Netherlands; 0.31 tonnes. Total weight in InterCatch; 419.3 tonnes. The remaining countries that did not upload data to InterCatch were France, Belgium, Portugal, Spain and parts of the UK; 91.4 tonnes in total.

Therefore, where relevant, the below sections are divided into two sections. The first details data from InterCatch; 419.3 tonnes) and the second (All data) details InterCatch data plus ICES preliminary landings ( 511 tonnes).

## Landings / catches

511 tonnes of saithe were landed in 2018 (All data; Table 31.1). This was a decrease of $31.8 \%$ compared to the 2017 ICES preliminary landings.

Ireland declared the highest catch (71\%) followed by the UK (16\%) and France (13\%). For the remaining countries; Belgium, Netherlands, Portugal and Spain, catches were negligible at $1 \%$.

## Catches by division

The predominant catches derived from Subarea 7 in 2018 ( $99.96 \%$ ), with a further $0.04 \%$ coming from Division 8.a (InterCatch data only). The majority of the catches were from Division 7.g ( $51.3 \%$ ), followed by $7 . j$ ( $42.2 \%$ ). The remaining divisions; $7 . \mathrm{a}-\mathrm{f}, \mathrm{h}$ and $8 . \mathrm{a}$ had very low catches and when combined constituted $6.5 \%$ of the catch.

## Catches by gear

The majority of saithe catches (InterCatch data only) originated from gillnets ( $71.6 \%$ ) followed by bottom trawlers (22.1\%), miscellaneous gears ( $2.9 \%$ ), seiners ( $2.5 \%$ ), trolling lines ( $0.5 \%$ ) and beam trawlers ( $0.4 \%$ ).

When the InterCatch data (419.3 tonnes) plus the unknown gear breakdown from the remaining preliminary data (91.4tonnes) were combined which was required for Table 6 of the advice sheet, the breakdown of catchesby gear type were gillnets ( $58.8 \%$ ) followed by bottom trawlers ( $18.2 \%$ ), miscellaneous gears ( $2.4 \%$ ), seiners ( $2.1 \%$ ), trolling lines ( $0.4 \%$ ), beam trawlers ( $0.4 \%$ ) and unknown (17.9\%).

## Catches by quarter

The breakdown of catch per quarter (InterCatch data only) shows that the highest landings were in quarter 2 ( $34 \%$ ) followed by quarter 1 (30\%), quarter 3 ( $17 \%$ ) and quarter 4 (19\%).

## Discards

A discard rate of $0.33 \%$ (1.4 tonnes) was reported by the Irish OTB fleet. No other discards were reported in InterCatch. Therefore, discarding is assumed negligible.

### 28.2 Stock assessment

A DCAC (Depleted-Corrected Adjusted Catch) method was used to estimate a yield likely to be sustainable (MacCall, 2009). TheDCAC-method was applied during WGCSE 2019 with the same model settings as was applied to pollack in Subarea 7 (ICES, 2018; 2019), due to the similarities of landings data. The saithe timeline begins in 1986; this is due to misreporting between saithe and pollack prior to 1986.

Nine separate model runs using various parameters (Table31.2) were conducted, producing an average DCAC value of 3152 tonnes plus upper ( 3596 tonnes) and lower ( 2555 tonnes) $95 \%$ confidence interval values.

The information provided for the assessment is insufficient to evaluate the exploitation and the trends of saithe in subareas 7-10. Commercial catches have generally declined, and in 2018 is amongst the lowest in the time-series.

## 2019 Results

The average DCAC values (Figure 31.2) show that landings are well below the average DCAC by 2656 tonnes. This suggests that yield could be increased up to 3152 tonnes in subareas 7-10.

### 28.2.1 Uncertainties in assessment and forecast

By construction, the DCAC method uses only long time-series of official landings and does not include any biological or survey data that are available for this stock.

The output is a smoothed value of the landings over the assessed time-series; therefore, the computations of DCAC will generate a similar value annually. Furthermore, the model cannot estimate reference points.

### 28.2.2 Recommendations

Saithe in subareas 7-10 is a new "stock" to WGCSE, and has never been assessed. Therefore, landings/catch data are the predominant source throughout the time-series.Supplementary data are available however, and shall be requested in the 2020 ICES datacall. WGCSE considers that pok.27.7-10 should be assessed using a different model therefore different assessment models shall be explored over the coming year with a purpose to Benchmark in 2021.

### 28.3 References

ICES. 2018. Advice basis. In Report of the ICES Advisory Committee, 2018. ICES Advice 2018, Book 1, Section 1.2. https://doi.org/10.17895/ices.pub. 4503.

ICES. 2019. Working Group for the Celtic Seas Ecoregion (WGCSE). ICES Scientific Reports. 1:29. xx pp. http://doi.org/10.17895/ices.pub. 4982.

MacCall, A. D. 2009. Depletion-corrected average catch: a simple formula for estimating sustainable yields in data-poor situations. ICES Journal of Marine Science, 66: 2267-2271.

Table 31.1. Landings of saithe in subareas 7-10 as officially reported to ICES.

|  |  |  |  | $\begin{aligned} & \text { त } \\ & \text { त्र } \\ & \text { 든 } \\ & \text { © } \end{aligned}$ | $\begin{aligned} & \text { ס } \\ & \underline{C} \\ & \underline{\pi} \end{aligned}$ |  |  | T0 0 0 0 0 |  | $\underset{ }{〕}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 25 | - | 8304 | - | 1739 | - | 40 | - | - | 1777 | 11885 | 0 | 11885 |
| 1987 | 20 | - | 6256 | - | 1624 | - | 2 | - | - | 2016 | 9918 | 0 | 9918 |
| 1988 | 24 | - | 6225 | 124 | 1400 | - | 1 | - | - | 1408 | 9182 | 0 | 9182 |
| 1989 | 16 | - | 8278 | 30 | 2165 | - | 16 | - | - | 1293 | 11798 | 0 | 11798 |
| 1990 | 9 | - | 6625 | - | 1068 | - | 24 | - | - | 2068 | 9794 | 0 | 9794 |
| 1991 | 5 | - | 7286 | - | 1495 | 1 | 29 | - | - | 2144 | 10960 | 0 | 10960 |
| 1992 | 2 | - | 1960 | - | 1721 | - | 38 | - | - | 1931 | 5652 | 0 | 5652 |
| 1993 | 4 | - | 1808 | - | 2010 | - | - | - | - | 2102 | 5924 | 0 | 5924 |
| 1994 | 9 | 1 | 3277 | - | 1915 | - | 7 | - | - | 2042 | 7251 | 0 | 7251 |
| 1995 | 8 | - | 2144 | - | 2382 |  | 14 | - | 13 | 1871 | 6432 | 0 | 6432 |
| 1996 | 5 | - | 2123 | - | 2062 | 3 | 13 | - | 27 | 2231 | 6464 | 0 | 6464 |
| 1997 | 9 | - | 1639 | - | 1384 | 2 | 7 | - | 23 | 1524 | 4588 | 0 | 4588 |
| 1998 | 8 | - | 1838 | - | 1431 | - | - | - | 68 | 983 | 4328 | 0 | 4328 |
| 1999 | 7 | - | - | - | 1352 | - | 5 | - | 35 | 716 | 2115 | 0 | 2115 |
| 2000 | 4 | - | 2720 | - | 1325 | - | 1 | - | 38 | 453 | 4541 | 0 | 4541 |
| 2001 | 7 | - | 911 | - | 1644 | - | 67 | - | 18 | 339 | 2986 | 0 | 2986 |
| 2002 | 13 | - | 578 | - | 1263 | - | 3 | - | 13 | 296 | 2166 | 0 | 2166 |
| 2003 | 3 | - | 457 | - | 754 | - |  | - | 9 | 366 | 1589 | 0 | 1589 |
| 2004 | 1 | - | 764 | - | 629 | - | 3 | - | 3 | 343 | 1743 | 0 | 1743 |
| 2005 | 1 | - | 396 | - | 394 | - | - | - | 21 | 210 | 1022 | 0 | 1022 |
| 2006 |  | - | 278 | - | 393 | - | - | 43 | 21 | 116 | 851 | 0 | 851 |
| 2007 | 1 | - | 326 | - | 286 | - | - | 3 | 4 | 87 | 707 | 0 | 707 |
| 2008 | 1 | - | 249 | - | 163 | - | - | 2 | 5 | 76 | 496 | 0 | 496 |
| 2009 | 1 | - | 231 | - | 254 | - | - | 3 | 3 | 112 | 604 | 0 | 604 |
| 2010 | 1 | - | 250 | - | 303 | - | - | 2 | 7 | 91 | 654 | 0 | 654 |
| 2011 | 1 | - | 229 | - | 685 | - | - | 4 | 10 | 69 | 998 | 0 | 998 |
| 2012 | 2 | - | 338 | - | 981 | - | 1 | 4 | 1 | 148 | 1475 | 0 | 1475 |
| 2013 | 2 | - | 269 | - | 1359 | - | - | 7 | - | 234 | 1871 | 0 | 1871 |
| 2014 | 2 | - | 117 | - | 1037 | 1 | - | 8 | 3 | 169 | 1337 | 0 | 1337 |
| 2015 | 1 | - | 93 | - | 659 | 1 | - | 5 | 1 | 102.2 | 862 | 0 | 862 |
| 2016 | <0.5 | - | 88 | - | 720 | $<0.5$ | - | 5 | 2 | 119.4 | 934 | 0 | 934 |
| 2017* | 1 | - | 94 | - | 588 | $<0.5$ | - | 4 | <0.5 | 62 | 749 | 0 | 749 |
| 2018* | 1 | - | 62 | - | 350 | <0.5 | - | 4 | <0.5 | 78 | 496 | 15 | 511 |

[^16]Table 31.2. Input parameters for the nine DCAC runs conducted for Saithe in subareas 7-10.

|  | Run 1 | Run 2 | Run 3 | Run 4 | Run 5 | Run 6 | Run 7 | Run 8 | Run 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| sumC | 132389 | 132389 | 132389 | 132389 | 132389 | 132389 | 132389 | 132389 | 132389 |
| CV sumC | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| no of years | 33 | 33 | 33 | 33 | 33 | 33 | 33 | 33 | 33 |
| iterations | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| stdev M | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 |
| $\mathrm{F}_{\text {MSY }} / \mathrm{M}$ | 0.6 | 0.8 | 1 | 0.6 | 0.8 | 1 | 0.6 | 0.8 | 1 |
| stdev $\mathrm{F}_{\text {MSY }}$ to M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| distr $\mathrm{F}_{\text {MSY }}$ to M | normal | normal | normal | normal | normal | normal | normal | normal | normal |
| $\mathrm{B}_{\mathrm{MSY}} / \mathrm{BO}$ | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 |
| stdev $\mathrm{B}_{\mathrm{MsV}} / \mathrm{BO}$ | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| up $\lim \mathrm{B}_{\mathrm{MSY}} / \mathrm{BO}$ | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| low lim $\mathrm{B}_{\text {MSY }} / \mathrm{BO}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| depletion delta $\Delta$ | 0.5 | 0.5 | 0.5 | 0.6 | 0.6 | 0.6 | 0.7 | 0.7 | 0.7 |
| stdev $\Delta$ | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| distr $\Delta$ | normal | normal | normal | normal | normal | normal | normal | normal | normal |



Figure 31.1. Saithe landings from WGNNSK and WGCSE from 1986-2018.

DCAC - Subareas 7-10


Figure 31.2. Saithe in subareas 7-10. The results of the depletion-corrected average catch (DCAC) assessment method as applied to commercial landings data since 1986. The grey box indicates the proxy for the maximum sustainable catch $\pm$ 95\% confidence intervals.

# 29 Seabass (Dicentrarchus labrax) in divisions 4.b-c, 7.a, and 7.d-h (central and southern North Sea, Irish Sea, English Channel, Bristol Channel, and Celtic Sea) 

## Type of assessment

This is an update of the assessment accepted as the agreed methods to use at the benchmark workshop for the seabass: WKBASS (ICES, 2017-2018). The assessment is performed using the Stock Synthesis model implementation (SS3; Methot, 2000; 2011). The stock is treated as Category 1 with a full analytical assessment and forecast.

## ICES advice applicable to 2019

The ICES advice for management of seabass fisheries in 2019 is available in the ICES Advice released in 2018, and states that "ICES advises that when the MSY approach is applied, total removals (includes commercial catch and recreational removals, taking mortality of released fish into account, estimated at approximately 5\%) in 2019 should be no more than 1789 tonnes.

### 29.1 General

### 29.1.1 Stock definition and ecosystem aspects

Studies including tagging programs, microchemistry and genetics are underway. These are designed to provide significant information on the movements of sea bass and could indicate the levels of mixing between stocks. Currently Atlantic stock identities are assumed to be as follows (ICES, 2012):


### 29.1.2 Management

Historical management is described in the Stock Annex.

### 29.1.2.1 Management applicable to $\mathbf{2 0 1 8}^{\mathbf{1}}$

Council Regulation (EL) 2018/120

1. It shall be prohibited for Union fishing vessels, as well as for any commercial fisheries from shore, to fish for European seabass in ICES divisions $4 b$ and $4 c$, and in ICES Subarea 7. It shall be prohibited to retain on board, tranship, relocate or land European seabass caught in that area.
2. By derogation from paragraph 1, in January 2018 and from 1 April to 31 December 2018, Union fishing vessels in ICES divisions $4 \mathrm{~b}, 4 \mathrm{c}, 7 \mathrm{~d}, 7 \mathrm{e}, 7 \mathrm{f}$ and 7 h and in waters within 12 nautical miles from baselines under the sovereignty of the United Kingdom in ICES divisions 7a and 7 g may fish for European seabass, and retain on board, tranship, relocate or land European seabass caught in that area with the following gears and within the following limits:
(a) using demersal trawls ${ }^{2}$, for unavoidable bycatches not exceeding 100 kilograms per month and $1 \%$ of the weight of the total catches of marine organisms on board caught by that vessel in any single day;
(b) using seines ${ }^{3}$, for unavoidable bycatches not exceeding 180 kilograms per month and $1 \%$ of the weight of the total catches of marine organisms on board caught by that vessel in any single day;
(c) using hooks and lines ${ }^{4}$, not exceeding 5 tonnes per vessel per year;
(d) using fixed gillnets ${ }^{5}$, for unavoidable bycatches not exceeding 1,2 tonnes per vessel per year. The derogations set out in the first subparagraph shall apply to Union fishing vessels that have recorded catches of European seabass over the period from 1 July 2015 to 30 September 2016: in point (c) with recorded catches using hooks and lines, and in point (d) with recorded catches using fixed gillnets. In the case of a replacement of a Union fishing vessel, Member States may allow the derogation to apply to another fishing vessel provided that the number of Union fishing vessels subject to the derogation and their overall fishing capacity do not increase.
3. The catch limits set in paragraph 2 shall not be transferable between vessels and, where a monthly limit applies, from one month to another. For Union fishing vessels using more than one gear in a single calendar month, the lowest catch limit set in paragraph 2 for either gear shall apply.

Member States shall report to the Commission all catches of European seabass per type of gear not later than 15 days after the end of each month.
4. In recreational fisheries, including from shore, in ICES divisions $4 \mathrm{~b}, 4 \mathrm{c}, 7 \mathrm{a}$ to 7 k , only catch-and-release fishing for European seabass shall be allowed. It shall be prohibited to retain on board, relocate, tranship or land European seabass caught in that area.

Council Regulation (EU) 2018/1308, Amended regulation (EU) 2018/120 as regards fishing opportunities for European seabass.

[^17]5. In recreational fisheries, including from shore:
(a) from 1 January 2018 to 30 September 2018, in ICES divisions 4.b, 4.c 7.a to 7.k, only catch-andrelease fishing for European seabass shall be allowed. During that period, it shall be prohibited to retain on board, relocate, transship or land European seabass caught in that area;
(b) from 1 October to 31 December 2018, in ICES divisions 4.b, 4.c 7.a to 7.k, not more than one specimen of European seabass may be retained per fisherman per day'.

### 29.1.3 Fishery description

### 29.1.3.1 Total landings (official)

The history of the fishery is described in the Stock Annex. Table 1 and Figure 1 presents official and total ICES landings. A large decrease in total landings was observed in the area in 2014 owing to poor weather conditions during winter and then from 2015 onwards owing to management measures put in place. Historically the bulk of the landings were made by the French fishery, but since implementation of management measures, landings are shared between French, UK and NL fisheries and to a lesser extent Belgium. In 2018, 912 tonnes were landed (official source): 431 t by UK, 297 t by France, 165 t by Netherland and 18 t by Belgium. Landings from France and UK by gear are given in Figure 2. In 2018, lines are the main gear used by both countries.

### 29.2 Data

### 29.2.1 Commercial landings

Landings series for use in the assessment are given in Table 2 for the six fleets for which selectivity is modelled: fleet 1- UK bottom trawls and nets; fleet 2- UK lines; fleet 3- UK midwater pair trawls; fleet 4- French combined fleets; fleet 5- other countries plus UK gears not included in fleet 1, with selectivity based on fleet 4; and fleet 6-recreational fisheries (2012 is the reference year with selectivity based on fleet 2, UK lines. The time-series of recreation fisheries is calculated iteratively so that fishing mortality is constant and equal to the fishing mortality in 2012 over the period 1985-2015) and with the implementation of the management measures a multiplier is applied from 2015 onwards (see chapter below). The landings figures are from census data (EU logbooks and/or sales slips) from several sources:

1. Official statistics recorded in the ICES official landings database since around the mid1970s (data from 1985 are used in the assessment);
2. French landings for 2000-2018 from a separate analysis by Ifremer of logbook, auction data and VMS data (SACROIS database);
3. Landings for Belgian vessels recorded in ICES database "InterCatch;"
4. Landings for Netherlands recorded in ICES database "InterCatch;"
5. UK landings by gear type recorded in official UK landings databases.

Details of the methodology used to calculate French and UK historical landings can be found in the Stock Annex.

### 29.2.2 Commercial length compositions

IBPBass 2016 developed the Stock Synthesis model to include both the length and age compositions for the landings of fleets for which selectivity is estimated (Fleet 1: UK combined trawl and nets -1985 onwards; Fleet 2: UK lines -1985 onwards; Fleet 3: UK midwater trawlers -1985 onwards; Fleet 4: French combined gears -2000 onwards). Fitting to length composition data helps
the estimation of length-based selectivity, whilst the age compositions (from application of agelength keys to length frequencies according to stratified sampling schemes) provide direct fitting of model estimates of catch-at-age. Since the length data are effectively being used twice, the length and age datasets are down-weighted (lambda values) to avoid over-fitting of the data. The composition data for the fleets are given in the SS3 data file. Input sample sizes for the multinomial composition data are derived from numbers of fishing trips sampled, as proxy for effective sample size. The relative sample sizes between years are maintained in any reweighting.

### 29.2.2.1 Sampling rates

$\mathrm{UK}(\mathrm{E} \& \mathrm{~W})$ sampling rates for age compositions, by gear group, are given in Table 3. Although ALKs are derived by the UK for separate sea areas, the same ALK is applied to all gear groups in an area meaning that the age composition estimates for the different gears are not independent. This was a principal motivating factor for IBPBass (ICES, 2014) to combine UK trawls, nets and lines into a single fleet for estimation of selectivity in Stock Synthesis.

The UK midwater trawl fleet landings were not sampled in 1997, 2013, 2014, 2015, 2016 and 2017 due to the small number of trips targeting bass. The UK at-sea sampling programme selects vessels at random from stratified vessel lists, which includes midwater pair trawlers in the same over 10 m vessel stratum as demersal otter trawlers, nets and lines. Similarly, port sampling is stratified by groups of ports, not métiers. The number of vessels and trips by midwater pair trawlers is very low, and therefore there is a high probability of low or zero numbers of samples. In Stock Synthesis, the missing age compositions for midwater trawls are imputed based on the selectivity parameters and the input landings figure. This has negligible impact on the assessment as this UK métier represented only $1 \%$ of total seabass landings in 2013 and landed only 1 t in 2014, less than 1 t in 2015, 2016, 2017 and 2018.

Sampling of seabass in France has also been very variable between areas and gears (Table 4). Numbers of trips are relatively stable from 2013, but numbers of fish sampled decreased from 2015 due to the implementation of management measures and the fact that relatively few fish are now landed.

NB: WGSCE 2018 was made aware of an issue with the sampling level in Q1 and Q2 of 2017 from France (The full explanation can be found in the working document from Quemar, Vigneau et al. "Estimation of quarterly length distribution of landings in the context of a six months disruption in the French on-shore sampling"). Because of the lack of market sampling for length (biological and on-board sampling was unaffected), efforts were made to try and fill the deficiency in the number of samples by use of simulation techniques. Both simulated data and actual data were uploaded to InterCatch combined making it impossible to distinguish true samples from those simulated. The simulation was based on commercial landings market categories. Thus for 2017, $4 \%$ of simulated samples have been implemented in the Bss 47 assessment model (which corresponds to $13 \%$ in term of fish measures).

Numbers of sampled trips for UK trawls, midwater trawls, nets and lines, and French all-gears, were used as proxies for effective sample size for initial development of the Stock Synthesis model for seabass by ICES IBPNew, IBPBass and IBPBass2.

Based on these results, the input effective sample sizes were iteratively adjusted using the Francis method of weighting, reducing the disproportionate effect of the different datasets used. The effective sample size which reflects the goodness-of-fit to the composition data are now fixed and additional data and associated sample sizes adjusted using the effective sample size multiplier for age and length compositions by fleet available in the stock annex.

### 29.2.2.2 Length composition estimates

Figure 3 and Table 5 gives fleet-raised length compositions for all French gears combined. French numbers-at-length are available from 2000 onwards. In the 2015 assessment (WGCSE 2015) a single fleet called "French fleet" was used. This fleet was the combination of several types of subfleets using various fishing gears: pelagic trawlers, bottom trawlers, netters, liners, Danish seiners and purse seiners. Figure 4 and Figure 5 give fleet-raised length compositions per UK métier used in the assessment (UK OTB-Nets; Lines; Midwater trawls).

### 29.2.3 Commercial age composition

Following to the IBPBass2 (2016) age compositions for French commercial fishery landings of seabass were used, derived from an annual age-length key (ALK) constructed for the whole area. It is applied to the total landings length frequency for the whole area (Table 6).
Fleet-raised age compositions were obtained for UK fleets from 1985 onwards by application of age-length keys developed for the areas $4 . b c, 7 . d, 7 . e$ and $h$, and $7 . a, f, g$. The annual age compositions for the combined otter and nets fleet and the line fleet are given in Table 7 and Table 8, and the age compositions for the UK midwater pair trawl fleet since 1996 are given in Table 9.

### 29.2.4 Commercial discards

Data sources for discards estimates and sampling design are described in the Stock Annex. Discarding of seabass by commercial fisheries can occur where:

- Fishing takes place in areas with bass smaller than the minimum landing size ( 36 cm in most European countries until 2015, then 42 cm ), and where mesh sizes $<100 \mathrm{~mm}$ are in use.
- Vessels catches are not in agreement with management measures.

Sampling rates and estimates of discards were provided to WGCSE from on-board sampling in the UK and France (Table 10, Table 11 and Table 12). The annual estimated quantities discarded by UK and French vessels from 2009 to 2014 has been less than 5\% of total landings. In 2016, 2017 and 2018 the level of discarding observed increased to $22 \%, 27 \%$ and $10 \%$ respectively from the French fleet in the area (Table 12). This was mainly attributed to bottom trawlers (for which seabass is often a bycatch). For the UK fleet a level of 9\% was observed in 2016 and $6 \%$ in 2018 mainly due to bottom trawlers. The level of available samples from the UK fleets in 2017 were too low to provide a raised estimate of total discards for this year for use in the model.

NB: Discards estimates from on-board sampling programme seems highly underestimated due to the low level of sampling on board: an extraction of French log book data in 2019 indicated that the level of discards estimated from on-board sampling is much lower. Two sources were used: French electronic logbooks ERSv1 and ERSv3 which progressively replaced ERSv1. The first does not spatialize the information which is reported for the whole fishing trip while the second does. Nevertheless, it is possible for ERSv1 to attribute discards to the stock $4 b c-7 d-h$ when the whole fishing trip is carried out in this area (which was the case for almost all the fishing trips observed). Using the different sources of data showed that from logbooks the French fisheries discard levels where in the region of 155.6, 270.9 and 456.4 tonnes as opposed to 152.7, 161.7 and 34.2 tonnes for 2016,2017 and 2018 respectively from the on-board sampling programme.

The increase in discards in 2018 may be explained by the increase of management measures but also by the fact that French fishermen have been encouraged to report their discards in logbooks with respect to the landings obligation regulation.

In the assessment French discarding values were revised for 2016 and 2017 and updated in 2018 with new estimates from French logbooks.

Most discards are fish below the minimum conservation reference size (MCRS) of 42 cm , and mostly from otter trawlers using $80-99 \mathrm{~mm}$ mesh in areas such as inshore regions of the English Channel where juvenile bass are most common.

There remains a large potential for bias in the discards estimates and it was recommended by England that the discards data for seabass should be used with caution. Particularly from under 10 m vessels, which take the bulk of the UK seabass catch, which was very infrequent until 2007, and line gears. However, mortality of discards for line gear was previously considered low. Research has recently been initiated on assessing the level of discards and post release mortality for this gear.

Previous assessments have excluded discards on the basis that the proportion discarded at an international level is relatively small ( $\sim 5 \%$ by weight). Discarding has been more of an issue recently since the inception of more restrictive management measures such as an increase in MCRS to 42 cm and bycatch limits for trawls and nets. The recent benchmarks WKBASS (ICES, 2017a; 2018) explored the performance of the Stock Synthesis model including recent (noisy) estimates of commercial discards and length compositions. For the years prior to the availability of discard observations from the observer schemes, a history of discards can be constructed based on the fishery selectivity and discarding ogives estimated for the recent years that have discards observations.

### 29.2.5 Recreational catches

IBPBass (2014) considered it necessary to have the catch and fishing mortality due to recreational fishing represented in the assessment model. The approach for achieving this has evolved since then through the benchmark process in 2016 (ICES, 2016) and updated during WKBASS in 20172018 (ICES, 2017b; 2018). The derivation of the recreational fishing catches is described in detail in Hyder et al. (2018), but the key points are summarised below along with some additional information on the generalisation of methods to account for any combination of season length and bag limit.

Two approaches are used to include recreational catches in the assessment:

- Before management measures (1985-2014): only a single year of recreational catches was available, so a constant recreational $F$ is applied in Stock Synthesis overall years. The F is derived from iteratively adjusting the assessment until the total retained recreational catch is equivalent to a value of $1440 t$ for a reference year of 2012. This value of $1440 t$ is obtained by summing international survey estimates for France, Netherlands and the UK obtained from surveys between 2009 and 2013. This represents total removals through summing the kept component and applying a $5 \%$ post-release mortality to the releases. A composite length-frequency distribution is generated for recreational removals based on survey data.
- After management measures (2015-present): given the management measures introduced for recreational fishers in 2015, it is unlikely that the assumption of constant $F$ is valid as release rates should increase. Limited survey data was available after the implementation of management, so it is not possible to use catch estimates. As a result, a method was developed for estimating the impact of combinations of the MCRS, season length and bag limits on removals by recreational fishing. This is used in both the assessment and forecast.


### 29.2.5.1 Recreational catches before management measures (1985-2014)

Survey data are available for France, UK and the Netherlands, but no survey data are available for Ireland, Belgium, Germany or Denmark (Table 13). An average of the two UK effort methods was included (Armstrong et al., 2013), French data was selected from the 2009-2011 study (Rocklin et al., 2014) and Netherlands data from 2010-2011 (van der Hammen and de Graaf, 2013) (Table 13). In addition, the original estimate of $60 t$ for Belgium was removed as the evidence underpinning this value was not available, but catches are likely to be low.

A study of post-release mortality of sea bass was combined with country-specific information on sea angling practices, the average post-release mortality of sea bass caught by recreational sea anglers in 2012 was $5.0 \%$ ( $95 \%$ CI=1.7-14.4\%) for BSS-47 (Lewin et al., 2018) (Section 27.2.6.7). Removals estimates were reworked for the 2012 reference year as the sum of retained fish and released fish with PRM of 5\% applied (Table 14). This gave a total removal of 1440 t for 2012 to be used within the assessment model (Table 14).

A single length composition for fishery removals was estimated for the stock based on the French and English length-frequency distributions from surveys (Armstrong et al., 2013; Rocklin et al., 2014). The raised length-frequency distributions for each country were binned into 2 cm lengths and summed for the kept and released components. Then a post-release mortality of $5 \%$ was applied to the released component before adding to the kept fish to give a total length-frequency distribution for the recreational fishery (Figure 8).

### 29.2.5.2 Recreational catches after management measures (2015-present)

The implementation of management measures should lead to a reduction in fishing mortality as more and larger fish are released. This means that it is not appropriate to assume constant recreational fishing mortality, so it is necessary to include an estimate of recreational catch or change in fishing mortality from 2015. However, coverage of surveys is patchy for all countries after 2015, with only provisional estimates available for the UK and the Netherlands. As a result, two potential methods are available for estimating catches or changes in fishing mortality:

1. Imputation: impute annual catches (kept and released) for England and France in 2016 by assuming the catches have changed over time to the same relative extent as Netherlands catch estimates between surveys in 2010-2011 or 2012-2013 and the survey in 20162017.
2. Reconstruction of change in recreational fishing mortality relative to the 2012 reference year: use the data from recreational surveys carried out by France, England, and Netherlands in 2009-2013 to calculate the reductions in retained catch in the observed trips if bag limits and increased MCRS had been implemented at the time of the surveys (Armstrong et al., 2014). The reductions in catch can be used to infer changes in recreational fishing mortality induced by changes in management, assuming full compliance and taking post-release mortality into account.

There are issues with both these methods. The use of imputation has a large uncertainty because: i) there are no time-series data to validate the assumption that national catches change to the same extent between years; ii) the surveys have sampling errors; and iii) the 2016-2017 Netherlands survey data are still provisional. The second method is also very uncertain due to sampling error and limitations in the survey data, assumptions concerning compliance, and dependence of results on the size of year classes present in the stock at the time of the surveys. However, the second method was considered more appropriate as it is based on observed data. As a result, the imputation approach was rejected, and estimation of the expected change in recreational F from in 2015 onwards due to change in MCRS, bag limits and closed seasons was carried out as described in Hyder et al. (2018). These reductions were used, along with post-release mortality of
$5 \%$, to calculate reductions in recreational F that may have occurred in 2015, 2016 and 2017 in response to the management measures, assuming full compliance (Table 15).

Full details of the methods and assumptions can be found in Hyder et al. (2018), but these have been extended to account for any length of season Management measures vary between areas both in terms of the measure implemented and the timing. For the BSS-47 stock, there was an increase to the MCRS to 42 cm and three fish bag limit for six months in 2015; an increase to the MCRS to 42 cm , six months no take, and a one fish bag limit for the remaining six months in 2016-2017; and an increase to the MCRS to 42 cm , nine months no take, and a one fish bag limit for the remaining three months in 2018 (Table 15).
To estimate the total removals (N_t) under different management scenarios, it is necessary to sum the numbers for each country ( $\mathrm{N}_{-}(\mathrm{t}, \mathrm{i})$ ) calculated from the numbers of retained fish ( $\mathrm{N} \_(\mathrm{h}, \mathrm{i})$ ), additional numbers dead releases of fish that would have been retained if no management were in place (N_(ar,i)), and the numbers of dead releases that would have occurred anyway (N_(or,i)), so:


If $p$ is the probability that a released fish dies and $r_{\_} i$ is the estimated reduction in retained fish in each country (i) under different management conditions (Table 16) (Armstrong et al., 2014) then the calculation of the 2012 equivalent numbers removed for each country under management applied for the whole year for any combination of bag limit and open season length (s) is:

$$
\begin{equation*}
\mathrm{N} \_(\mathrm{t}, \mathrm{i})=\mathrm{s}\left(1-\mathrm{r} \_\mathrm{i}\right) \mathrm{N} \_(\mathrm{h}, \mathrm{i})+\mathrm{p}\left(1-\mathrm{s}+\mathrm{sr} \_\mathrm{i}\right) \mathrm{N} \_(\mathrm{h}, \mathrm{i})+\mathrm{pN} \mathrm{~N}_{-}(\mathrm{r}, \mathrm{i}) . \tag{2}
\end{equation*}
$$

This applies to the management measures for 2016-17 and 2018, but not for 2015 as the emergency measures were implemented part way through the year. As a result, the reduction in numbers under the 2015 management measures is:

$$
\begin{equation*}
N_{-}(t, i)=\left(1-r_{-} i 2\right) N_{-}(h, i)+\left(p r \_i N_{-}(h, i)\right) 2+p N_{-}(r, i) . \tag{3}
\end{equation*}
$$

For each management scenario, summing across countries gives total recreational removals in numbers that would have been expected in the years of the surveys. The ratio of numbers removed in each scenario to the removals with no management can then be used to infer reductions in recreational fishing mortality in the years when the management measures came into force, for use in the stock assessment. These reductions in fishing mortality are only approximate as the contribution of year classes in the years of the surveys will be different to the composition of catches in the years when management was changed. The reductions in recreational fishing mortality are unlikely to be fully realised due to non-compliance and if post-release mortality is greater than $5 \%$ on average. The reductions in terms of numbers implied a potential F multiplier for existing management measures (Table 15) and any combination of management measures (Table 17).

### 29.2.5.3 Inclusion in the stock synthesis model

For the period 1985-2014 before management measures were introduced, recreational catch was iteratively reconstructed conditioned following previous assessments (ICES 2016; 2018) on the 2012 estimated value of 1440 t (Table 14). The selectivity was based on length-frequency distributions of removals assuming a $5 \%$ post-release mortality (Figure 8). Management measures were introduced in 2015, including an increase in the MCRS and various combinations of season length and bag limits. A multiplier was derived from 2012 catches in terms of numbers of fish for the recreational $F$ that related to the reduction in catch due to management. Frec multipliers were applied of 0.832 in 2015, 0.282 in 2016 and 2017, and 0.191 in 2018 (Table 15). Frec multipliers
were also calculated for different combinations of bag limit and season length for inclusion in the forecast (Table 17).

### 29.2.6 Biological data

This section provides biological parameters of growth, maturity and natural mortality required for the stock assessment of seabass. Further information and plots of growth and maturity data can be found in the stock annex and WGCSE 2013, and detailed methods and results are given in IBPNew 2012 (ICES, 2012a) working documents by Armstrong (2012) and Armstrong and Walmsley (2012b,c). Further information of natural mortality data can be found in the WKBASS report (ICES, 2017b).

### 29.2.6.1 Growth parameters

Growth parameters, standard deviations of length-at-age distributions, and an age error vector are input to the Stock Synthesis model. These are derived from more than 90000 seabass sampled by Cefas since 1985 from fishery catches around England and Wales as well as from trawls surveys of young bass in the Solent and Thames estuary.

The sampled seabass shows some sexual dimorphism of growth from about seven years of age onwards. It is currently not possible to implement a sex-disaggregated Stock Synthesis assessment as it is impossible to disaggregate commercial fishery catches and survey catches by sex. Therefore, a combined-sex assessment using a combined-sex growth curve is adopted. Mean length-at-age has not shown any trend over time, and length-at-age is also very similar in strong and weak year classes (Armstrong and Walmsley, 2012b). Hence data have been combined over the full series to estimate growth parameters, and the estimated body weights-at-length and age in the Stock Synthesis assessment model are treated as being constant over time.

Von Bertalanffy model parameters were estimated by area using an absolute error model minimizing $\sum(\text { obs-exp })^{\wedge} 2$ ) in lengths-at-age:

| Area | 4.bc | 7.d | 7.e | 7.afg | All areas |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Linf (cm) | 82.98 | 87.22 | 92.27 | 81.87 | 84.55 |
| K | 0.1104 | 0.09298 | 0.07697 | 0.09246 | 0.09699 |
| t0 (years) | -0.608 | -0.592 | -1.693 | -1.066 | -0.730 |

The "all areas" VBGF parameters are used in the Stock Synthesis model.

### 29.2.6.2 Standard deviations of length-at-age

As expected, the standard deviation of length-at-age increased with length, and the trend could be described by the linear model $\mathrm{SD}=0.1166^{*}$ age +3.5609 . The regression estimates of SD by age class are input to the assessment model to generate length-at-age distributions.

### 29.2.6.3 Age error parameters for Stock Synthesis

Inclusion of age error parameters in the Stock Synthesis model (CVs for ageing error by age class) were derived from results of the ICES seabass scale exchange in 2002 (Mahé et al., 2012). CVs of $12 \%$ at-age were specified as increasing values per age class to give a standard error of $\sim 1$ year per age class. These are used in the SS3 observation submodel to derive expected values for observed data on age distributions. Further information on ageing precision and calibration between laboratories will become available from an ICES calibration study in 2015.

### 29.2.6.4 Weight-at-length

Weight-at-length and age was estimated within the Stock Synthesis model according to the following relationship derived from UK sampling:
$\mathrm{W}(\mathrm{kg})=0.00001296 \mathrm{~L}(\mathrm{~cm})^{\wedge} 2.969$

### 29.2.6.5 Maturity-at-length

Maturity ogives are derived from 590 male and 730 female seabass collected in the UK between 1982 and 2009 immediately prior to and during the spawning season (December to April). The data were modelled using a binomial error structure and logit link function, fitted in R to individual observations (Armstrong and Walmsley, 2012b). The logistic model describing proportion mature by 1 cm length class $L$ was formulated as:
$\operatorname{Pmat}(\mathrm{L})=1 /(1+\mathrm{e}-(\mathrm{a}+\mathrm{bL}))$
defined by the parameters slope $b$ and length intercept $a$. These parameters were estimated separately for females and males. This can also be expressed as:
$\operatorname{Pmat}(\mathrm{L})=1 /(1+e-b(L+c))$ where $\mathrm{c}=\mathrm{a} / \mathrm{b}$
Stock Synthesis uses the second formulation, and the parameters required are the slope $(b=$ 0.3335: entered as a negative value) and the length inflection, which is the estimated length at $50 \%$ maturity $(\mathrm{L} 50 \%=40.65 \mathrm{~cm})$.

The parameters of the model $\operatorname{Pmat}(\mathrm{L})=1 /(1+e-b(L+c))$ are given in Table 18.
The logistic model for females and males is:
$\operatorname{Pmat}(\mathrm{L})=1 /(1+\mathrm{e}-0.3335(\mathrm{~L}-40.6488)) \quad$ (females)
$\operatorname{Pmat}(\mathrm{L})=1 /(1+\mathrm{e}-0.4861(\mathrm{~L}-34.6652)) \quad$ (males)

The length-based maturity ogive for female seabass is used in the current Stock Synthesis assessment model, which derives proportion mature at age by applying the length-based ogive to the length-at-age distributions defined by the growth parameters and SD of length-at-age (Table 19).

### 29.2.6.6 Natural mortality

The current assessment uses a value of $M=0.24$ for all ages and years. This was derived based on methods using information on longevity, growth and maturity. The maximum observed age (tmax) in over 90000 age readings in the UK since the 1980s was 28 years. Data from 1145 recreationally caught seabass Ireland reported by IBPNew (ICES, 2012) showed a maximum age of 26. The Hoenig (1983) method based only on maximum age for teleosts gave M of $0.15-0.16$ for maximum age 26-28 (Table 20). A more recent paper by Then et al. (2015) analysed data from 226 studies (including Hoenig, 1993) to evaluate the robustness of life-history based M inferences. They propose maximum age methods as being the most robust. Their equation $M=4.899$. tmax${ }^{0.916}$ gives M values of $0.23-0.25$ for tmax of $28-26$ years (Table 31.2.8.6.1). They also give an expression using values of von Bertalanffy parameters $K$ and $\operatorname{Linf}(M=4.118 . K 0.73$.Linf-0.33) which predicts $\mathrm{M}=0.17$ for seabass in areas 4 and 7 . The WKBASS Data WK proposed the use of the Then et al. tmax method $(M=0.24)$ as being more robust than inferences from any single study.

Natural mortality is high in young fish and declines with age, as shown by multispecies models that include diet data and estimation of size preferences (such as applied by ICES WGSAM for the North Sea, ICES, 2014). Proxy methods to infer age-dependent M in younger fish are given by Lorenzen (1996) and Gislason et al. (2010). Values for seabass by age are given in Tables 20 and 21 and Figure 12. The Gislason method gives much lower M for adult fish. Brodziak et al. (2009) suggest that methods such as Lorenzen can be used to derive the relative age-dependent
patterns for younger fish, but can be re-scaled to give $M$ at older ages more similar to those from methods using (e.g.) tmax. Table 21 and Figure 12 show Lorenzen and Gislason M rescaled to give mean M at ages 10 and older that are equivalent to the Then et al. (2015) prediction of 0.24 for tmax 26-28, or rescaled to the previous M value of 0.15 . The WKBASS Data WK proposed Lorenzen scaled to 0.24 for the older ages. For the benchmark, the following M options were explored:

1. $\mathrm{M}=0.15$ at all ages (continuity with previous approach);
2. $\quad M=0.24$ at all ages, Then et al., 2015;
3. Lorenzen scaled to $\mathrm{M}=0.24$ at ages $10+$;
4. Estimation of age-invariant M by model;
$\mathrm{M}=0.1$ and 0.2 at all ages to explore likelihoods in comparison with the other options.
There are no direct estimates of M for seabass. The WKBASS 2017 Data WK reviewed a number of life-history based methods for inferring natural mortality rates in teleost fish based on metrics such as lifespan and growth parameters. The WKBASS 2017 assessment WK adopted the predictions from a recent paper by Then et al. (2015) which analysed data from 226 studies to evaluate the robustness of life-history based M inferences. Their equation $\mathrm{M}=4.899$. tmax ${ }^{-0.916}$ gives M values of 0.23-0.25 for tmax of 26-28 years as observed in the BSS-47 stock. WKBASS 2017 Data WK also considered methods to derive age-dependent M (Gislason et al., 2010; Lorenzen, 1996) and to rescale these to match the Then et al. (2015) prediction over the age range of mature fish. However this was not adopted for the benchmark assessment which adopted $M=0.24$ for all age groups.

### 29.2.6.7 Post-release mortality

## Commercial fisheries

Discarding of seabass below the MCRS occurs in most commercial fisheries to a variable extent. Previously, ICES advice sheets indicate overall international discard rates of only $5 \%$ by weight for the BSS.27.4bc7ad-h stock based on data supplied to the Working Group on the Celtic Seas Ecoregion (WGCSE). The WGCSE and WKBASS Data WK 2017 showed that discard rates have typically been highest in bottom otter trawls (OTB) and have increased following the introduction of additional management measures in 2015. Discards are now included in the assessment of this stock and in the absence of any data on post-release survival, this has been assumed to be zero for all commercial fisheries. This will overestimate commercial fishing mortality to some extent although the effect will be small due to the low discard rates.

Survival of fish discarded by commercial line vessels may be similar to survival of recreational angling releases (see next section), but work is needed to establish the typical gear, handling, and condition of fish to be released. Survival of seabass caught by trawls, seines, fixed or driftnets and longlines will depend on many factors including tow duration, soaking times, gear design, deep-hooking, and time on deck. WKBASS identified a need for studies on post-release survival of seabass in different commercial fisheries, particularly in view of the potential inclusion in the Landings Obligation.

## Recreational fisheries

Releases of can be mandatory or voluntary and can represent a large proportion (>50\%) of the catch, so must be accounted for in the assessment (Ferter et al., 2013). Post-release mortality of hook-and-line caught fish is not easy to measure and can vary significantly between species and fisheries. Many factors are also important including water temperature, hooking damage, and handling, so it is important to account for different fishing practices. Recreational fisheries on

European seabass are characterised by relatively high release rates, which appear to have increased following changes in management which increased the MCRS from 36 cm to 42 cm in 2015 and imposed bag limits and closed seasons.

Existing studies of post-release mortality of seabass and other similar species were reviewed (ICES, 2017b). Based on the information provided by Hyder et al. (2018), WKBASS agreed on a figure of $5 \%$ for PRM in recreational fisheries on BSS.27.4bc7ad-h, which are predominantly sea angling (ICES, 2018). This estimate is based on Lewin et al. (2018) in which 144 fish were maintained in an aquaculture facility and then captured by experimental angling using a range of bait and artificial lures. The fish were then released, and held for ten days to assess mortality. The effects of different bait types, air exposure, and deep hooking were investigated, with increased mortality associated with use of natural bait ( $13.9 \%, 95 \% \mathrm{CI}=4.7-29.5 \%$ ) and deep hooking ( $76.5 \%, 95 \% \mathrm{CI}=50.0-93.2 \%$ ). By combining the experimental results with country-specific information on sea angling practices, the average post-release mortality of seabass caught by recreational sea anglers in 2012 was $5.0 \%$ ( $95 \% \mathrm{CI}=1.7-14.4 \%$ ) for BSS-47 (Lewin et al., 2018).

### 29.2.7 Survey data used in assessment

### 29.2.7.1 Pre-recruit surveys in UK

An inshore trawl survey in autumn in the vicinity of a major bass nursery area in the Solent (7.d English coast, Figure 9) provides abundance indices at-ages 2 to 4 for the stock assessment. Data are available from 1982, although there are intermittent years when the survey did not take place (Table 22). The stock annex provides details of this survey and of some other pre-recruit survey series not considered appropriate by previous WGs and IBPBass for inclusion in the assessment. A previous assessment of the stock by Pawson et al. (2007), using a statistical separable model, indicated that recruitment patterns in 7.afg, 7.eh, 7.d and 4.bc were similar to the trends in the Solent survey. This provides some justification for using the Solent survey in the current assessment despite its extremely localised coverage.

Abundance indices for ages 2-4 in the Solent autumn survey have large interannual variability (Table 22; Figure 10). Strong year classes are apparent in 1989, 1995 and 1997, but in the last decade, year-class strength has been less variable, a pattern also seen in the commercial fishery. The survey indicates a general trend of increasing recruitment since the early 1990s, but weak year classes from 2008 to 2012. There is only one pronounced year-effect, in 2007. The age- 2 index appears less consistent than the age 3 and 4 indices.

### 29.2.7.2 Pre-recruit surveys in France

Similar surveys, carried out by Ifremer, began on the coast of France from 2014. At this stage, the methodology has been set and is starting to produce additional information on seabass age groups $0,1,2,3$ providing an early insight into the French coast nursery dynamics and the beginnings of a new time-series. In the Channel, the survey takes place in the Seine estuary and preliminary indices are available from 2017. The survey will continue until 2021 under a European Maritime and Fisheries Fund (EMFF) program (NOURDEM).

### 29.2.7.3 Channel Groundfish survey CGFS

The Ifremer Channel Groundfish survey, is carried out in October each year since 1988, provides swept-area indices of seabass abundance in the Eastern Channel (7.d) together with length compositions. Details of the survey are given in Coppin et al. (2002), which includes a full description of the GOV trawl, sampling stations are shown in Figure 11. The majority of seabass are caught in the coastal waters of England and France (Figure 11).

The abundance indices are calculated applying a stratified-random swept-area based estimator. Strata correspond to ICES statistical rectangles. Swept-area is calculated using wingspread. As
this is a stratified swept-area based indicator, uncertainty is based on between haul variance within a strata and summation of variances across strata. Further details are in the stock annex.

The swept-area indices are given in Table 23. The trends in both the index and in the proportion of stations with seabass showed similarities to the trend in total biomass estimates from the ICES, WGCSE 2013 update assessment using Stock Synthesis, before the survey data became available, lending a priori support to the use of the index in the assessment (Figure 12).

The precision of the swept-area indices appears unrealistically high in some years (e.g. 0.025 in 1991), which may indicate that the index trends are driven largely by the incidence of positive catches. Modelling of the data using delta lognormal models may provide more realistic precision. During trial Stock Synthesis runs, the use of the CVs in Table 23 resulted in poor fit to length frequencies in many years due to individual years with very low CVs being given far too much weight. Relaxing the CVs to 0.30 for all years except the first three years (set to 0.6 in preliminary runs given the very low incidence of positive stations) allowed the model to fit the length compositions more closely over the series. The final assessment excluded the composition data for 1988-1990 due to the very low sample sizes, but retained the overall index.

NB: The Channel Ground Fish Survey (CGFS) has been conducted since 1988 with a systematic fixed sampling program with a high opening (GOV) bottom trawl ( 20 mm mesh size codend), using the same Research Vessel Gwen Drez since 1988 to 2015. The RV Gwen Drez was decommissioned in 2015, and survey continue the time-series using the RV Thalassa (a bigger vessel). An inter-calibration exercise was conducted in 2015 by using paired tows, simultaneously with both vessels (see Working Document in WGIBTS 2015 report for description of the inter-calibration results). The original index was calculated as numbers of fish per hour tow. The initial step in calculating the index was numbers per ICES square per hour tow (the stratum in this survey) and then raised to the whole Eastern Channel to compute a number of fish per age class per hour tow. As the surface trawled area differed between the two RVs (difference in trawling speed and width of the gear used) a density index (number of fish per $\mathrm{km}^{2}$ ) was also calculated in order to create a consistent index over the whole time-series. The index is then computed using the formula:

$$
\bar{N}=\frac{\sum_{s} A_{s} \overline{N_{s}}}{\sum_{s} A_{s}} \quad \begin{aligned}
& \overline{N_{s}} \text { mean abundance in the strata } s, \text { expressed in number } / \mathrm{km}^{2}
\end{aligned}
$$

With :

As the vertical opening of the gear used by the RV Thalassa was higher than the previous one, and in order to take into account any vessel effect on catchability, the CPUE were compared for all the species caught. Differences in CPUEs between the new and the old survey setting were found for nine species (mostly pelagic species) and a correction factor applied to continue the time-series. The correction coefficient for seabass used to continue the time-series is $\mathrm{R}=1.707+/-$ 0.091. In addition to the calculation of the new index, a number of errors were found in the surface calculation of some strata. These errors where corrected and the new indices (expressed in number of fish per $\mathrm{km}^{2}$ instead of number of fish per hour fished) take these corrections into account.

As there were significant changes in how the index was calculated along with the introduction of a new vessel and gear the EWG agreed that a full review was needed before the index could be used in the assessment. The review revealed that the index could be used up to and including 2014 and the series from 2015 onward should be considered a new time-series. Therefore the CGFS index discontinues in 2014.

The new time-series from 2020 when five years of data become available.

### 29.2.8 Commercial landings per unit of effort

Following the recommendation from WKBASS 2018 the French LPUE index is now calculated by modelling the zeros and non-zeros values using a delta-GLM approach (see stock Annex for details). In 2019 a revision to the calculation considering the level fixed to define zeros and nonzeros values used was carried out. It was considered that an error in the threshold to define positive values (used in WGCSE 2018) for modelling positive proportion was incorrect. The threshold should be defined as 0.99 kg and not 0.05 kg (Figure 13). Confidence interval calculated through a bootstrap estimation are presented in Figure 14 with the agreed LPUE series used in the final assessment. Further details of the LPUE update are in WD 06. The outcome of this update to the LPUE has resulted in a change of perception of the stock, see subsequent sections.

### 29.2.9 Other relevant data

None.

### 29.3 Stock assessment

### 29.3.1 Model structure and input data / parameters for update assessment

The assessment was conducted using Stock Synthesis (Methot, 2000; 2011), using version 3.24u (Methot, 2011). The structure and input data / parameters of the SS3 model are summarized below and details are available in previous sections:

### 29.3.1.1 Model structure

- Temporal unit: annual based data (landings, discards, survey and commercial tuning indices, age- and length-frequencies);
- Spatial structure: One area;
- Sex: Both sexes combined.


### 29.3.1.2 Fleet definition

Six fleets defined: 1. UK bottom trawls, nets; 2. UK lines; 3. UK midwater trawls; 4 . French fleets (combined); 5. Other (other countries and other UK fleets combined); 6. Recreational fisheries.

### 29.3.1.3 Landed catches

Annual landings in tonnes from 1985 to final year for the Five fleets from ICES subdivisions 4.b and c, 7.a, d-h. French data were as provided by Ifremer and the recreational catch was provide for 2012 with the time-series from 1985 to present iteratively reconstructed conditioned on the 2012 estimated value of 1440 t .

### 29.3.1.4 Abundance indices

Channel Groundfish Survey in 7.d in autumn (France), 1988 to 2014: total swept-area abundance index and associated length composition data (Table 23). Number of stations with seabass is used as input effective sample size. Input CV for survey $=0.60$ for 1988-1990 and 0.30 for 1991 to 2014. First three years of composition data are excluded due to sampling levels and higher uncertainty in the data.

Cefas Solent Autumn bass survey (7.d), years 1986 to 2009, 2011, 2013 to present, for ages 2-4. Selection was fitted as a function of length using a double normal model, with minimum and maximum ages specified as 2 and 4 in the age selection function (Table 24).

### 29.3.1.5 Fishery landings age composition data

Age bins: 0 to 15 with a plus group for ages 16 and over. Age compositions for fleets are expressed as fleet-raised numbers-at-age, although they are treated as relative compositions in SS3. Year range for UK trawls/nets and UK lines: 1985 to present; UK midwater pair trawl: 1996 to 2018 (no samples for 1997, 2013-2014, 2016-2017); French all fleets combined were input from 2000 to present.

### 29.3.1.6 Fishery landings length composition data

The length bin was set from 4 to 100 cm by 2 cm intervals. Length compositions for fleets are expressed as fleet-raised number-at-length. Year range for UK trawls/nets: 1985 to present; UK lines: 1985 to present; UK midwater pair trawl: 1985 to 2012 (no samples for 1997, 2013-2018); French all fleets combined were input from 2000 to present.

### 29.3.1.7 Model assumptions and parameters

Table 25 summarises key model assumptions and parameters. Other parameter values and input data characteristics are defined in the SS3 control file BassIVVII.ctl, the start.SS file, the forecast file Forecast.SS and the data file BassIVVII.dat.

### 29.3.1.8 Incorporation of recreational fishery catch estimates

Where catch is not available the vector of recreational fishing catch values, landed plus dead released fish assumed to be $5 \%$ of all released fish, was generated using the selectivity from the 2012 length compositions obtained from the recreation survey and a value of F for recreational fishing in 2012. For a given value of F, the recreational harvest was calculated based on catch in 2012 and the recreational F. The F and landings for recreational fishing was adjusted in successive SS3 runs until the recreational F for the time-series was close to the F giving 1440 t in 2012. An F multiplier was applied to the Fs after 2014 to take account of the new management measures. The calculations for the final assessment run are given in Table 28.

### 29.3.1.9 Final update assessment: diagnostics

The likelihood components ( $\log L^{*}$ Lambda) for the update SS3 assessment are given below:

| Likelihood components | Likelihood |
| :--- | :---: |
| TOTAL | 705.0 |
| Catch | $5.38 \mathrm{e}-13$ |
| Equilibrium catch | 0.0289 |
| Survey | -45.3 |
| Discards | 21.82 |
| Age compositions compositions | 379.7 |
| Recruitment | 325.5 |
| Parameter soft bounds | 23.26 |
| Convergence level | $1.89 \mathrm{e}-7$ |

A range of model outputs and diagnostics are given in Figures 15-33.
Good correspondence was found between the observed and fitted length and age compositions for each fleet (Figures 20-29), although the fit to the French length compositions in 2018 was poorer than for preceding years. Some diagonal residual patterns are noted in the commercial age compositions indicating some problems in fitting extreme variations in recruitment.

Any smearing of age estimates from a strong year class into neighbouring weak ones could be responsible for year-class residuals in the UK age compositions that are apparent in the first half of the series. The age error vector included in the model helps to accommodate this in the fit to age compositions. The combined fit of the age and length composition data aggregated over the series was very close (Figure 23 and 29).

The survey abundance indices both fisheries-independent and fishery-dependent are fit reasonably well (Figure 30 to 32). The UK Solent autumn survey is characterised by a large variability with outliers present in the model fit (Figure 30). The model fits closely to the low indices for recent years because there are few fishery composition data for estimating these recent year classes.

The model is able to fit recruitment deviations with reasonable precision back to around the 1974 year class (Figure 33) allowing a longer term perception of recruitment dynamics. Recruitment is highly variable with no evidence of a reduction in average recruitment at the lower SSB values (Figure 33) although this perception is affected by the imposition of a steepness value of 0.999 for the fitted Beverton-Holt stock-recruit curve. Sensitives to differing values for this parameter carried out during the benchmark workshops found that likelihoods progressively worsened as the steepness value was reduced.

### 29.3.2 Analytical retrospective analyses

Retrospective analysis with a seven-year peel was carried out and a five-year peel for the calculation of the Mohn's rho. This analysis shows that there is some evidence of a retrospective pattern, see table below and Figure 34, for recruitment, SSB and fishing mortality. However, the retrospective bias is within the tolerance threshold accepted by ICES $(-15$ to +20$)$ for SSB and fishing mortality, there has been no tolerance threshold set for recruitment.

|  | Mohn's rho |
| :--- | :---: |
| Spawn-stock biomass | 0.152 |
| Fishing mortality (ages 4-15) | -0.125 |
| Recruitment (age 0) | 0.796 |

The model is sensitive to the recent change in selectivity due to management measures where a block change in the selectivity and retention parameter estimates where introduced for data proceeding 2015. The model is also sensitive to the addition of new data and its associated weightings with the recent data having more influence given the higher sampling levels.

### 29.3.3 Final update assessment: long-term trends

The time-series of estimates of numbers-at-age, combined recreational and commercial $\mathrm{F}(4-15)$, are given in Tables 26 and 27, and a summary of SSB, recruitment, $F$ and commercial and recreational catch are given in Table 28 and Figure 35. These series are based on the final SS3 update run with 2018 set as the final year. In order to obtain biomass estimates for 2019 and Fs for 2018 for the forecast the final year is set to 2019.
A sharp increase in F between 2011 and 2013 is generated because the assessment model interprets that landings were maintained despite a rapid decline in biomass. This may be a plausible scenario where aggregations or predictable migration routes of seabass can be targeted, and it is possible for fisheries to maintain landings as total stock size declines, and hence inflict an increasing fishing mortality rate. The F has since decreased in line with sharp reduction in catches due to the discontinuation of the French midwater trawl and the implementation of additional management measures.

WGCSE concludes that strong year classes in 1989 and some subsequent years caused a rapid increase in biomass throughout the stock area, and landings and fishing mortality in the commercial fishery also increased. The combined commercial and recreational fishery F has been well above the Fmsy proxies estimated during WKBASS and the recent update of reference points due to the correction to the LPUE series. Recruitment has been declining since the mid-2000s, and has been very poor since 2008, however the recruitment estimated for 2013 and 2014 is above the long-term geometric mean of 16018 . Uncertainties in the assessment are explored in a subsequent section.

### 29.3.4 Comparison with previous assessments

With the addition of the 2018 data and new LPUE series, the time-series of recreational catch was updated to remain consistent with the assumption of a constant F for the period 1985 to 2014 and an F multiplier reduction for 2015 to present (Figure 36).

With these changes included in the update assessment the perception of the stock has changed with spawning-stock biomass showing a higher biomass level and lower $F$ than that estimated in 2018 for the recent period.

The spawning-stock biomass, fishing mortality and recruitment estimated in 2018 when compared with the recent assessment (Figure 37) is within the $95 \%$ confidence intervals. However, as the stock is at its lowest level of the time-series, around Blim, has high uncertainty and a retrospective bias the EWG agreed that reference points would need to be updated with the WGCSE 2019 accepted assessment. The outcome of the analysis to update reference points is discussed in a subsequent section.

### 29.3.5 The state of the stock

The marked increase in biomass in the 1990s was driven by the very strong 1989 year class and a number of subsequent year classes. The biomass prior to this was declining during a period of poor recruitment, and the recent decline in biomass also coincides with a period of poor recruitment, but under conditions of higher F than estimated for the 1980s. The stock has been characterised by periods of poor recruitment in the 1980s and now again since 2008. These periods of poor recruitment have a major impact on biomass, which is exacerbated by any increase in F. Total biomass reacts more quickly than SSB due to the delayed maturity.

The period of increasing SSB in the 1990s and early 2000s also coincided with expansion of the stock in the North Sea. The enhanced productivity and geographic range of the stock at this time also coincided with a period of elevated sea temperatures (see WGCSE and stock annex for UK inshore sea temperature trends in relation to seabass recruitment).

The assumption of a constant recreational fishing mortality over time implies that recreational harvests were a much larger fraction of total fishery removals in the 1980s compared with the 2000s onwards (Figure 15). It is likely that in the 1970s or earlier, seabass were primarily the target of recreational fishing.

### 29.4 Biological reference points

The fishing pressure and biomass PA and MSY reference points defined by WKBASS 2018 were updated using the WGCSE 2019 assessment due to the inclusion of additional 2018 data and a new LPUE series which changed the perception of the stock.

WGCSE noted that due to the changes in perception of the stock biomass and fishing pressure and how close the stock is to Blim the reference points would need to be updated to reflect these amendments. Sensitivities to the new input data and an update of reference points using the most recent accepted assessment was carried out and presented in a working document (WD 03). The same method as that carried out by WKBass 2018 using the EqSim R packages was implemented with the only change being that the time-series of selectivity used was increased from two to four years as additional data were available post-implementation of new management measures.

The following table summarises the updated reference points using the most recent accepted assessment:

| Stock | Seabass in ICES divisions 4.b,c and 7.a, d-h. |  |  |
| :---: | :---: | :---: | :---: |
| PA Reference points | Value 2019 | Rational |  |
| $\mathrm{Bl}_{\text {lim }}$ | 10313 | Lowest observed SSB (Type 5 S-R relationship) |  |
| $\mathrm{B}_{\mathrm{pa}}$ | 14439 | $\mathrm{B}_{\mathrm{lim}} \times 1.4$ |  |
| $\mathrm{F}_{\text {lim }}$ | 0.254 | In equilibrium gives a $50 \%$ probability of $\mathrm{SSB}>\mathrm{B}_{\text {lim }}$ |  |
| $\mathrm{F}_{\mathrm{pa}}$ | 0.1815 | $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} / 1.4$ |  |
| MSY Reference point |  |  |  |
| $\mathrm{F}_{\text {MSY }}$ | 0.1713 (0.197) | Reduce as $\mathrm{F}_{\mathrm{MSY}}>\mathrm{F}_{\text {PA }}>\mathrm{F}_{\mathrm{P} .05}$ |  |
| $\mathrm{F}_{\text {MSY lower }}$ | 0.142 | Reduce as $\mathrm{F}_{\mathrm{MSY}}>\mathrm{F}_{\mathrm{PA}}>\mathrm{F}_{\mathrm{P} .05}$ |  |
| $\mathrm{F}_{\text {MSY upper }}$ | 0.1713 | Reduce as $\mathrm{F}_{\mathrm{MSY}}>\mathrm{F}_{\mathrm{PA}}>\mathrm{F}_{\mathrm{P} .05}$ | $\underset{\substack{\circ}}{\stackrel{\leftrightharpoons}{7}}$ |
| MSY $\mathrm{B}_{\text {trigger }}$ | 14439 | $\mathrm{B}_{\mathrm{pa}}$ |  |

### 29.5 Short-term predictions

Inputs for a short-term forecast are given in Table 29, and their derivation is explained below.

### 29.5.1 Recruiting year-class strength

Recruitment estimates for seabass are below average from 2008 to 2012 (Table 28). SS3 does not estimate recruit deviations for years with no survey data for that year class. Hence, the model imputes a value from the stock-recruit curve at virgin biomass for year classes 2017 and after. This value (22 031 thousand) differs slightly from the 1985 to 2013 geometric mean (16 018 thousand) which was adopted for subsequent year classes for the forecast. However, this year the working group agreed that recruitment is at a low level and has been since 2005 and agreed to only include 2005 to 2016 for the geometric mean recruitment for the forecast (12 383 thousand). This is summarised in the text table below:

| Year class | SS3 (age 0) | GM 2005-2016 |
| :--- | :--- | :--- |
| 2016 | 13452 thousand |  |
| 2017 | 12383 thousand |  |
| 2018 | 12383 thousand |  |
| 2019 | 12383 thousand |  |

WGCSE (2013 and 2014) reviewed some information on environmental influences on seabass recruitment which supports the apparent recent reduction in recruitment from 2008-2012. Survival of 0-gp and 1-gp seabass in nursery areas in estuaries and salt marshes is thought to be enhanced by warmer conditions promoting survival through the first two winters, and increasing the growth rates (Pawson, 1992). Data on coastal sea temperatures in the south of the UK were presented by WGCSE to show that shifts between periods of poor recruitment and periods of above-average recruitment were associated with changes from cooler to warmer sea conditions, and that recent poor recruitment from 2008 onwards coincided with cooler conditions (see stock annex). During 2014, sea temperatures off southern England were exceptionally warm, which may have favoured survival and growth of young bass. The Solent survey in 2017 indicated that numbers of 2-gp bass (2015-year class) was below the average of the time-series. Although the evidence is weak, it is not a critical assumption for short-term forecasts as this year class has very little impact on the short-term forecast.

### 29.5.2 Numbers of fish in 2019

These were derived from the update Stock Synthesis run with final year set at 2018. The numbers for ages 0-2 in 2019 were adjusted using the ratio of LTGM to SS3 values for 2017-2019 age 0 as explained above.

### 29.5.3 F-at-age vectors

Status quo F-at-age for the commercial fishery was taken as the average F-at-age as estimated from the last three years scaled to the final year derived from the update Stock Synthesis run with final year set at 2018. This approach was taken to allow for the change in selectivity associated with the implementation of new management measures (Figure 16b).

The recreational F vector was estimated in a similar way using the average of the last three years, however the final $\mathrm{F}_{\text {bar }}$ was scaled using F multipliers on the 2012 F in Table 17 taking into account the management measures in place. For 2018, this was a three-month open season with a one bag limit and a MCRS of 42 cm . Additional years Fs were scaled taking into consideration management measures in place for 2019 and where management measures are unknown, a partial F proportion based on pre management measures was applied up to the maximum that the F multiplier would allow, 0.663 . The F multiplier of 0.663 takes into account the increase in the MCRS to 42 cm .

### 29.5.4 Weights-at-age

Mean weights-at-age in the stock were taken from the Stock Synthesis output. The commercial fishery weights for 2018 were derived as a weighted mean of the values for French and UK fleets given in the Stock Synthesis output, using the model estimates of catch numbers for the two fleets as weighting factors. The annual weights-at-age for any fleet are time-invariant, as they are derived from length-at-age derived from von Bertalanffy growth curve parameters, with selectivity applied where appropriate. Length at AMAX (30 years) was estimated as 84.12 cm .

### 29.5.5 Maturity ogive

The proportion mature at-age is the length-based ogive applied to the length-at-age distributions around the input VB growth curve, calculated within Stock Synthesis.

### 29.5.6 Detailed short-term forecast output at status quo $F$

A detailed short-term forecast is given in Table 30 assuming that F in 2019 and 2020 is the average of the previous three years scaled to 2018 from the assessment for the commercial fleet for the recreation fleet the partial $F$ used is that described in Section 27.5.3.

Fishing at the same fishing mortality as in 2018 for the commercial fleets and a one bag limit for seven months for the recreation fishery will result in a further decline in SSB from 11413 t in 2020 to 11338 t in 2021 remaining just above the $B_{\lim }$ of 10313 t .

There is uncertainty in the forecast, as the actual rate of decline in population abundance in recent years is likely to be more uncertain than indicated by the SS3 model confidence limits. Also, the final package of technical and other management measures for seabass in 2015 to present are not fully known at this stage, and information will be needed on their implementation and effectiveness before their impact on fishing mortality can be ascertained. The assumption of constant recreational F is also untested.

### 29.5.7 Management options

WGCSE provides management options in which F multipliers are applied proportionally to commercial and recreational F-at-age (Table 31). In reality, management may wish to allocate the combined forecasted landings in any way considered appropriate, and this would imply differing F-multipliers applied to each fishery.

The management options table includes options for F multipliers for a number of different scenarios and include the multiplier giving the proposed $\mathrm{Fmsy}^{\text {m of }} 0.1713$ for combined commercial and recreational fishing. With zero F in 2020, SSB is expected to increase from 11413 t in 2020 to 12371 t in 2021. At FMSY, the combined commercial and recreational catch in 2019 is expected to be around 2427 t . However, as SSB is predicted to be below MSY Btrigger in 2020 and 2021 FmSY is adjusted accordingly and expected catches are thus reduced to 1939 t . When compared with estimated landings for all fisheries of 1248 t in 2019, this represents an eight percent increase in combined commercial and recreational catches. The allocation between commercial and recreational fisheries depends on the balance of controls applied on recreational and commercial fishing in 2019 and 2020.

### 29.6 Uncertainties and bias in assessment and forecast

### 29.6.1 Landings and discards data

The historical fishery catch data are subject to several biases. From 2000 to 2015, French landings data from the ICES commercial landings database are replaced by more accurate figures from a separate analysis of logbook, auction data and VMS. From 2011 onwards, the official and scientific French landings use the same analysis of logbook and auction data and VMS data. Prior to 2000 official French landings figures have had to be redistributed between ICES areas, according to the average spatial pattern observed from 2000 onwards.

Historical landings of small-scale national fisheries not supplying EU logbooks or sales slips are known to be inaccurate. IBPBass ran the Stock Synthesis model with and without additional UK landings for nets and lines estimated from a separate Cefas logbook scheme, and found this had relatively little impact on stock trends or fishing mortality, but rescaled the biomass and recruitment due to the additional catch. However, if the extent of non-reporting is changing over time,
for example to develop track record in the possible event of a future TAC, then bias will be introduced in the assessment trends.

Discard rates are low in most fisheries other than trawls. Estimates of discards are available only from the early 2000s, but do not cover all fisheries, are imprecise, and are only included for some fleets in the assessment. The overall discard rate by weight is thought to be less than $5 \%$ before the implementation of management measures increasing in recent years. Nonetheless, a timeseries of discards at-length or -age is needed for all fleets if the impact of technical measures to improve selectivity is to be evaluated as part of any future bass management.

### 29.6.2 Fishery composition data

The ability to fit selectivity patterns for defined groups of fishery métiers, and to detect changes in selectivity, depends crucially on collection of adequate numbers of independent, representative samples of length and age to sufficiently characterise the length or age compositions of the selected métier groups. What constitutes "sufficient" is impossible to define without simulation studies to examine relationship between precision of input data and the precision of estimates required for management.

The absence of length composition data for French fisheries prior to 2000 is a serious deficiency in the model preventing any evaluation of changes in selectivity that may have occurred, for example due to changes in the mix of gear types. The numbers of trips of each métier group sampled on shore in France and the UK has varied widely over time, and in the UK has declined substantially since the 2000s. Currently there are no composition data supplied by Netherlands and Belgium.

ICES has developed extensive advice on establishing statistically-sound sampling designs for estimating fishery length and age compositions and discard quantities (see reports of ICES Workshops on Practical Implementation of Statistically Sound Catch Sampling Programmes (WKPICS1-3, available on ICES website). Stratified random sampling of fishing vessels or harbours may lead to low sample sizes for species such as seabass for which large fractions of the total catches may be taken in relatively small numbers of fishing trips. The cost-benefit of expanding the sampling in vessel or harbour strata where most seabass landings are recorded, without compromising statistical sampling design, should be investigated. The next benchmark should evaluate if sampling is currently sufficient to support continued application of Stock Synthesis fitting selection parameters to fishery composition data.

The comparative assessment using age compositions for French fleets showed that these data may improve the robustness of the assessment in future, and this should be subject to an interbenchmark assessment and peer review.

### 29.6.3 Recreational fishery harvests

Current assessments accommodate an estimate of recreational fishery landings in the assessment and forecasts based on landing from 2012 (ICES, 2016; 2018). This a crude approach based on surveys for only a year or two in France, UK, and the Netherlands, and leads to an assumption of constant recreational fishing mortality over time. Recreational catches have been observed to vary significantly over time in other fisheries, so it is possible that this assumption of constant mortality is unlikely to be true. For example, the estimate of recreational harvest in the Netherlands has varied over time. In addition, no data are currently available after the implementation of management measures, meaning that a reconstruction procedure is needed to estimate the impact of management on recreational catches. Further survey data are needed to confirm the
level of recreational catches and releases, and to develop a time-series to evaluate changes in recreational fishing mortality and any changes in selectivity.

Release rates are expected to increase due to bag limits and increases in MLS that are in place or planned. Current studies of post-release mortality are limited, and more studies are needed to develop a better understanding of the fate of released fish given the high incidence of catch-andrelease practices in sea angling for seabass. WGCSE must collaborate closely with the ICES Working Group on Recreational Fishery Surveys to identify priorities for future surveys and hooking mortality studies.

A full set of data for the UK, France, Netherlands, and Belgium should be available by 2020, and it is possible that additional information may be available on post-release mortality. As a result, recreational removals should be reviewed and updated once these survey data become available.

### 29.6.4 Surveys

The Channel Groundfish Survey surveys included in the assessment provides data on a wider range of sizes and ages than the Cefas Solent survey, though with a steeply domed size selection pattern. From 2015 onwards, Ifremer no longer use the scientific vessel "Gwen Drez" which will be replaced by the larger vessel "Thalassa". A calibration exercise was carried out in 2014 to assess the effect of this change to a larger vessel. WGCSE noted a concern that coverage of the coastal waters of $7 . \mathrm{d}$ could be altered by the use of this new vessel (the size of the vessel may prevent fishing as close to the coast as was possible with the previous vessel). The results of the calibration exercise where evaluated and it was found that the series could not be extended beyond 2014 and that a new series would need to be created from 2015 onward.

The Cefas pre-recruit surveys are now reduced to just the Solent autumn survey, with the Solent spring and the Thames survey having been removed by previous benchmark assessments as being unsuitable. Recruitment estimates for the most recent years are heavily dependent on the Solent survey, and it is important to maintain this series. However, there is a need for information on recruitment trends in other areas, as it cannot be assumed that the Solent index will in the long term represent overall recruitment patterns throughout areas 4 and 7. A study by France under the EU Framework for Community actions in the field of water policy (Table 32) shows clearly that seabass nurseries in the Channel have asynchronous patterns of abundance of young bass. In the UK, 37 seabass nursery areas such as estuaries and saltmarshes are defined for implementing conservation measures, and there are others that may be added. Similar habitats for young bass also occur in France and the Netherlands. A more robust survey design would treat individual nursery grounds as strata or station clusters in an internationally coordinated, stratified survey design. The possibility for this, and the sampling effort and costs for a desired precision, could be considered as part of a long-term seabass management plan.

### 29.6.5 Commercial Ipue indices

The reliance of the assessment on the Solent and Channel trawl surveys is a potential source of bias because they cover only a part of the stock range, and the selectivity is heavily skewed towards young bass. This is of principle concern in establishing the current rate of decline in spawning-stock biomass and associated trends in fishing mortality. In the absence of relative abundance indices for older bass from surveys or commercial fishing vessels covering the range of the stock, it is difficult for the model to fix the recent stock trends and fishing mortality. Statistical modelling of French LPUE data by vessel and rectangle by Laurec and Drogou (WGCSE 2015, Annex 3, WD 07) is now used in the assessment. In parallel a study on effect of vessel selection has been completed (Bissery, Mahevas and Drogou), but is still under development and cannot be evaluated yet.

Analyses of UK commercial fishery LPUE, based on averaging across ICES rectangles where the bulk of seabass catches have been recorded, was presented to IBPNEW in 2012. There were divergent trends between fleets where seabass are typically a bycatch, and mainly under 10 m vessels where increased targeting has probably been occurring using lines and nets. Future development of UK LPUE indices together with equivalent French data would require careful evaluation of potential for LPUE of each fleet to track abundance.

### 29.6.6 Stock structure and migrations

The assessment treats all seabass in 4.b,c and 7.a,d-h as a single biological stock. Although there can be extensive migrations, for example between the south of the area and the Bay of Biscay (which is treated separately in the WGBIE group), or between the North Sea and the Channel, there is also strong site fidelity (Pawson et al., 2008) resulting in a high proportion of tagged fish being recaptured at the same coastal location, even in subsequent years after migrations to offshore spawning sites. Immature seabass may remain close inshore, and exploitation of young fish in coastal waters ( $<6$ nautical miles offshore) may be predominantly by inshore fleets of that country. Mature fish originating from coastal waters of the UK, France or Netherlands or other countries may become increasingly vulnerable to offshore pelagic pair trawlers fishing mainly on mature fish during December to April. These spatial, ontogenetic patterns may lead to complex responses of length and age compositions to previous fishery catches of each country and fleet. This could potentially be addressed using spatial structuring in Stock Synthesis, but the data demands would increase substantially. Both the UK (England) and France have embarked on major programmes of bass research involving electronic and conventional tagging, and modelling of larval drift patterns, to try and improve knowledge of spatial dynamics.

### 29.6.7 Biological parameters

The maturity ogive used in the assessment was derived from sampling from the 1980s onwards. There has been no coordinated sampling across the full range of the stock in recent years to determine if the current ogive is still valid. Sporadic recent sampling has suggested that seabass may be spawning at sizes smaller than recorded historically (see stock annex). This would alter the Fmsy and could also be associated with changes in growth parameters. Mean length-at-age in UK samples remained more or less constant over several decades of sampling, but this analysis needs updating. Changes in growth, or inappropriate growth parameters, will lead to bias in fitting length-selectivity parameters to the French fishery and survey data.

### 29.6.8 Intermediate year fishing mortality and catch levels for forecasts

As the Measures introduced by the EU commission to reduce fishing mortality toward FmsY, have the potential to affect the short-term forecast assumptions for this stock. The working group agreed that the fishing mortality in 2019 under the same measures are likely to be similar to those in 2018. In the absence of any data on changes in selectivity, there would be no reason to deviate much from this assumption.

Tables 30 provides a detailed short-term status quo forecast and Table 31 provide a range of management options from the forecast run.

### 29.7 Recommendations

### 29.7.1 Management considerations

Seabass in this stock are characterised by slow growth, late maturity and low natural mortality on adults, which imply the need for comparatively low rates of fishing mortality to avoid depletion of spawning potential in each year class. Productivity of the stock is affected by extended periods of enhanced or reduced recruitment, which appear to be related to changes in sea temperature. Warm conditions facilitate northward penetration of seabass in the North Sea and Northeast Atlantic and enhance the growth and survival of young fish in estuarine and other coastal nursery habitats. A period of above-average sea temperatures and enhanced recruitment between 1989 and the mid-2000s generated a large increase in biomass and a geographic expansion. Increased abundance and a lack of a TAC or other means to control fishing outside of nursery areas stimulated a growth of fisheries and markets for seabass. Many small-scale artisanal fisheries, especially line fishing and some forms of netting, have developed a high seasonal dependency on seabass, and there is also a significant recreational fishing mortality in inshore waters. The behaviour of bass, forming predictable aggregations for spawning and moving close inshore to feed at other times of year, increase their vulnerability to exploitation by offshore and inshore fisheries. Increased targeting of seabass has resulted in a progressive increase in fishing mortality above values considered appropriate to achieve Fmsy. The combination of increasing fishing mortality and environmental conditions causing poor recruitment since 2008 appears responsible for a continuous decline in biomass since 2010. Catches appear to be declining in fisheries where seabass is mainly a bycatch, but some other fisheries such as netting in the UK appear to be expanding and may be exploiting known seasonal migration routes and local aggregations of fish despite a more widespread contraction of the population.

Careful management of fishing pressure on seabass is needed to prevent SSB declining to such an extent that the stock's ability to produce strong recruitment in more favourable environmental conditions is impaired. Since 2013, the European Commission has been in dialogue with Member States to develop a package of management measures to promote recovery of the stock. This resulted in emergency measures to stop the offshore pelagic trawl fishery on spawning aggregations between January and April 2015, bag limits for recreational fishing, and proposals to increase the MLS to 42 cm . Further measures to restrict catches without resorting to a TAC have been implemented. Any management measures applied to commercial and recreational fisheries should take into account the need for collection of data to demonstrate the effectiveness of the measures, and the ability to enforce the measures adequately.

ICES advice in 2004 recommended that "implementation of 'input' controls, preferably through technical measures aimed at protecting juvenile fish, in conjunction with entry limitations into the offshore fishery in particular should be promoted", and that "any consideration of catch limitation (output control) would need to take into account that seabass are a bycatch in mixed fisheries to a various extent, depending on gear and country; this incites discarding and should be avoided". This form of advice has re-occurred in subsequent ICES advice for seabass.

WGCSE notes that protection of juvenile fish through technical measures is good to improve the fishery selectivity and increase the number of seabass that are able to spawn at least once, but this is probably not enough to ensure a sufficient decrease in F. Protection of juveniles already exists to an extent already through designation of 37 UK seabass nursery areas where certain types of fishing on seabass is prevented annually or seasonally. However, catching and discarding of seabass by trawlers fishing close to nursery areas remains an issue. Data available to WGCSE indicate that discarding is mainly by otter trawlers using $80-90 \mathrm{~mm}$ mesh in or near areas where juvenile bass are most abundant, for example in UK coastal waters of the eastern

Channel. Improvements to fishery selectivity to successfully achieve a large reduction in fishing mortality on pre-spawning fish without increasing discarding would require changes to gear designs, which could have a strong spatial management component.

Entry limitation can prevent an increase in effort but will not decrease $F$ to the extent needed, unless existing licences are withdrawn. The occurrence of seabass as a small bycatch in many fisheries raises the problem of this becoming a "choke species" if vessel catch limits are introduced under EU legislation and seabass fall under the landings obligation.

ICES also previously advised that "Management of seabass fisheries needs to take into account the distinctive characteristics and economic value of the different fisheries. Seabass is of high social and economic value to the large inshore artisanal fleets and to sea angling and other recreational fishing that contribute substantially to local economies". Data from France indicate that the first sale value of the high-volume and lower quality catches of seabass caught by pelagic trawlers targeting offshore spawning fish during December to March has been up to three times lower per kg than for smaller volume sales of higher-quality fish for métiers fishing inshore (Drogou et al., 2011). However, there is at present insufficient information to accurately evaluate the total economic value and impact of seabass fisheries beyond just the first sale value and covering direct incomes from sales and direct as well as indirect and induced costs, employment and added value generated downstream. The interrelationship between markets for wild caught and farmed seabass also needs to be evaluated. A number of studies on the economic value of recreational sea fisheries have been conducted in recent years, and these demonstrate high levels of spend into national economies; for example the total direct, indirect and induced spend of sea angling in England in 2012 was estimated at $£ 2$ bn GBP (Armstrong et al., 2013) although this cannot be easily allocated to a spend per species.

No bio-economic scenarios are available at present to appreciate the effect of management measures for seabass, based on economic considerations, and work is urgently needed in this area. The importance of seabass to recreational fisheries, artisanal and other inshore commercial fisheries and large-scale offshore fisheries in different regions means that resource sharing is an important management consideration that has implications for the type of scientific evidence needed. WGCSE has estimated that up to 30\% of total landings in France, England, the Netherlands and Belgium were attributable to the recreational fisheries in recent years.

The effects of targeting of offshore spawning aggregations of seabass in the English Channel and Celtic Sea are poorly understood, particularly how the fishing effort is distributed in relation to mixing of fish from different nursery grounds or summer feeding grounds in the UK, France and other countries, given the strong site fidelity of seabass. This is a subject of a new scientific study on seabass in the UK.

The current stock structure assumptions are pragmatic and need further evaluation. The seabass population in coastal waters of the Republic of Ireland is currently considered as a separate stock, although it extends into at least one of the ICES divisions defining the 4.bc and 7.a,d-h stock. Further studies are needed to determine if the seabass in Irish coastal waters are indeed functionally separate, or if they also mix with the other stock during spawning time and contribute to commercial catches on the offshore spawning grounds. Moreover, the Bay of Biscay is also currently considered as a separate stock although tagging programme indicates some exchange with the area 4 and 7 stock studied assessed by WGCSE.

As bass is, at present, a non-TAC species, there is potential for continued displacement of fishing effort from other species with limiting quotas. The effort of the pelagic fisheries during winter and spring can shift between the Bay of Biscay and the English Channel and approaches, and there is evidence for such a shift to the Channel in recent years, which is likely to have increased the fishing mortality on seabass in Area 7. The fisheries on seabass have grown in the 1990s and 2000s due to good recruitment, and new markets have been established, competing with farmed
bass. Fishing mortality has gradually increased over time and has been above Fmsy for many years. With the stock in decline and no effective control on these fisheries, the risk of stock collapse is currently very high unless strong year classes are produced again. Therefore, in addition to technical measures to improve the fishery selection pattern, an overall limitation of total fishing mortality across all ages of seabass needs to be continued through appropriate measures.

### 29.8 References

Armstrong, M.J. 2012. Life-history estimates of natural mortality of seabass around the UK. Working Document: ICES IBPNew 2012; October 2012. 3 pp.

Armstrong, M., and Walmsley S. 2012b. Age and growth of seabass sampled around the UK. Working Document: ICES IBPNew 2012; October 2012. 15 pp.

Armstrong, M., and Walmsley, S. 2012c. Maturity of seabass sampled around the UK. Working Document: ICES IBPNew 2012; October 2012. 14 pp.

Armstrong, M., and Maxwell, D. 2012. Commercial fleet lpue trends for seabass around the UK. Working Document: ICES IBPNew 2012; October 2012. 29 pp.

Armstrong, M., Le Goff, R., and van der Hammen, T. 2014. Assessment of recreational fisheries for seabass. Request for Services - Sea bass. Commitment No.686192. Paper for Scientific, Technical \& Economic Committee for Fisheries of the European Commission, Brussels, Belgium. 36 pp.

Armstrong, M., Brown, A., Hargreaves, J., Hyder, K., Munday, M., Proctor, S., and Roberts, A. 2013. Sea Angling 2012 - a survey of recreational sea angling activity and economic value in England. Crown copyright 2013, London, UK. 16 pp.

Coppin, F., Le Roy, D., Schlaich, Y. 2002. Manuel des protocoles de campagne halieutique: Campagnes CGFS, Système d'information halieutiques - Campagnes à la mer. Ifremer, 09/2001. DRV/RH/DT/ANNUMERO, 29 pp. (in French).

Drogou, M., Biseau, A., Berthou, P., de Pontual, H., Habasque, J and le Grand, C. 2011. Synthèse des informations disponibles sur le Bar: flottilles, captures, marché. Reflexions autour de mesures degestion. http://archimer.ifremer.fr/doc/00035/14577/11879.pdf.

Ferter, K., Weltersbach, M. S., Strehlow, H. V., Volstad, J. H., Alos, J., Arlinghaus, R., Armstrong, M., et al. 2013b. Unexpectedly high catch-and-release rates in European marine recreational fisheries: implications for science and management. ICES Journal of Marine Science, 70: 1319-1329.

Hyder, K., Readdy, L., Drogou, M., Woillez, M., and Armstrong, M. 2018. Recreational catches, post-release mortality and selectivity. WKBASS working document. 18 pp.

ICES. 1998. Report of the study group on the precautionary approach to fisheries management. ICES Copenhagen, 3-6 February 1998. ICES C.M. 1998/Assess:10.

ICES. 2003. Study Group on Precautionary Reference Points for Advice on Fishery Management. Copenhagen, 24-26 February 2003. ICES CM 2003/ACFM:15.

ICES. 2008. Report of the Working Group on the Assessment of New MoU Species (WGNEW). By correspondence, ICES CM 2008/ACOM:25. 77 pp .

ICES. 2012a. Report of the Inter-Benchmark Protocol on New Species (Turbot and Seabass; IBPNew 2012). ICES CM. 2012/ACOM:45.

ICES. 2012b. Report of the Working Group on Assessment of New MoU Species (WGNEW), 5-9 March 2012, ICES CM 2012/ACOM:20. 258 pp.

ICES. 2012c. Report of the Working Group on Recreational Fisheries Surveys (WGRFS). ICES CM 2012/ACOM:23. 55 pp .

ICES. 2014. Report of the Working Group for Celtic Seas Ecoregion (WGCSE), 8-17 May 2013, Copenhagen, Denmark. ICES CM 2013/ACOM:12. 1986 pp.

ICES. 2014b. Report of the Inter-Benchmark Protocol for Seabass in the Irish Sea, Celtic Sea, English Channel, and Southern North Sea (IBPBass). By correspondence. ICES CM 2014/ ACOM:46.

ICES. 2014c. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 13-22 May 2014, Copenhagen, Denmark. ICES CM 2014/ACOM:12.
ICES. 2016. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 4-13 May 2016, Copenhagen, Denmark. ICES CM 2016/ACOM:13. 1312 pp.

ICES. 2017a. Report of the Working Group on Recreational Fisheries Surveys (WGRFS), 12-16 June 2017, Horta, Azores, Portugal. ICES CM 2017/EOSG:20. 113 pp.
ICES. 2017b. Report of the Data Evaluation meeting for the Benchmark Workshop on Sea Bass (DEWKBASS), 10-12 January 2017, Copenhagen, Denmark. ICES CM 2017/ACOM:32. 139 pp.
ICES. 2018. Report of the Benchmark Workshop on Seabass (WKBASS), 20-24 February 2017 and 21-23 February 2018, Copenhagen, Denmark. ICES CM 2018/ACOM:44. 287 pp.

Lewin, W.C., Strehlow, H.V., Ferter, K., Hyder, K., Niemax, J., Herrmann, J.-P., and Weltersbach, M.S. 2018. Estimating post-release mortality of European seabass based on experimental angling. ICES J. Mar. Sci. doi:10.1093/icesjms/fsx240/.

Mahé, K., Holmes, A., Huet, J., Sévin, K., Elleboode, R. 2012. Report of the Seabass (Dicentrachus labrax) Otolith and Scale Exchange Scheme 2011, 16 pp.

Methot, R.D. 2000. Technical Description of the Stock Synthesis Assessment Program. National Marine Fisheries Service, Seattle, WA. NOAA Tech Memo. NMFS-NWFSC-43: 46 pp.
Methot, R.D. 2011. User Manual for Stock Synthesis, Model Version 3.23b. NOAA Fisheries Service, Seattle. 167 pp.

Pawson, M. G. 1992. Climatic influences on the spawning success, growth and recruitment of bass (Dicentrarchus labrax L.) in British Waters. ICES mar. Science Symp. 195: 388-392.
Pawson, M. G., Kupschus, S. and Pickett, G. D. 2007. The status of seabass (Dicentrarchus labrax) stocks around England and Wales, derived using a separable catch-at-age model, and implications for fisheries management. ICES Journal of Marine Science 64, 346-356.

Pawson, M.G., Brown, M., Leballeur, J. and Pickett, G.D. 2008. Will philopatry in seabass, Dicentrarchus labrax, facilitate the use of catch-restricted areas for management of recreational fisheries? Fisheries Research 93 (2008) 240-243.

Quemar T., Vigneau J., Dubroca L. 2018. Estimation of quarterly length distribution of landings in the context of a 6 -months disruption in the French on-shore sampling. Working Document to ICES/WGBIE 2018.

Quirijns, F. and Bierman, S. 2012. Growth and maturity of seabass sampled around the Netherlands. Working Document: ICES IBPNew 2012; October 2012. 9 pp.
Then, A. Y., J. M. Hoenig, N. G. Hall, and D. A. Hewitt. 2015. Evaluating the predictive performance of empirical estimators of natural mortality rate using information on over 200 fish species. ICES Journal of Marine Science 72:82-92.

Van der Hammen, T and de Graaf, M. 2012. Recreational fishery in the Netherlands: demographics and catch estimates in marine and freshwater. IMARES Wageningen UR, Report Number C147/13.
Van der Hammen T, and de Graaf, M. 2015. Recreational fisheries in the Netherlands: analyses of the 20122013 logbook survey, 2013 online screening survey and 2013 random digit dialing survey. IMARES C042/15, pp 55.

Table 1. Bss.27.4.bc, 7.a, d-h: Annual landings from 4.b,c and 7.d,e-h (Official landings per country and Total Ices estimates).
$\left.\begin{array}{lcccccccccc}\hline \text { Year } & \text { Belgium } & \text { Denmark } & \text { Germany } & \text { France* } & \text { UK } & \begin{array}{l}\text { Nether } \\ \text { lands }\end{array} & \begin{array}{l}\text { Channel } \\ \text { Is. }\end{array} & \begin{array}{l}\text { Total offi- } \\ \text { cial }\end{array} & \text { Total } \\ \text { ICES }\end{array}\right]$

Source: Official Landings Statistics. 2018 provisional data; Total ICES, from InterCatch database.

Table 2. Bss.27.4.bc, 7.a, d-h. Landings for the country / fleet components included separately in the assessment model.

| Year | Fleet 1 : UK <br> Trawls, nets | Fleet 2 : UK Lines | Fleet 3 : UK pelagic trawlers | Fleet 4 : France combined gears | Fleet 5: Other countries and gears | Fleet 6 : RecFish |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 70 | 30 | 1 | 870 | 23 | 2148 |
| 1986 | 84 | 33 | 2 | 1180 | 19 | 1933 |
| 1987 | 96 | 18 | 0 | 1840 | 25 | 1753 |
| 1988 | 129 | 30 | 8 | 1028 | 44 | 1616 |
| 1989 | 141 | 29 | 7 | 917 | 67 | 1490 |
| 1990 | 128 | 18 | 22 | 849 | 47 | 1342 |
| 1991 | 152 | 60 | 14 | 971 | 29 | 1224 |
| 1992 | 105 | 23 | 8 | 1001 | 49 | 1222 |
| 1993 | 146 | 62 | 1 | 979 | 68 | 1383 |
| 1994 | 354 | 154 | 0 | 786 | 76 | 1640 |
| 1995 | 424 | 169 | 4 | 1057 | 181 | 1848 |
| 1996 | 308 | 128 | 87 | 2395 | 104 | 1890 |
| 1997 | 335 | 119 | 71 | 1984 | 111 | 1819 |
| 1998 | 241 | 121 | 85 | 1773 | 170 | 1766 |
| 1999 | 274 | 148 | 220 | 1843 | 185 | 1765 |
| 2000 | 236 | 53 | 52 | 1805 | 261 | 1816 |
| 2001 | 263 | 58 | 97 | 1883 | 199 | 1898 |
| 2002 | 361 | 75 | 110 | 1825 | 251 | 1980 |
| 2003 | 353 | 65 | 127 | 2471 | 443 | 2035 |
| 2004 | 380 | 72 | 131 | 2604 | 544 | 2048 |
| 2005 | 353 | 59 | 68 | 3161 | 789 | 2014 |
| 2006 | 359 | 119 | 11 | 3259 | 629 | 1955 |
| 2007 | 413 | 166 | 37 | 2771 | 677 | 1922 |
| 2008 | 514 | 163 | 17 | 2750 | 663 | 1902 |
| 2009 | 486 | 147 | 9 | 2649 | 598 | 1859 |
| 2010 | 452 | 183 | 42 | 3236 | 649 | 1751 |
| 2011 | 462 | 143 | 98 | 2526 | 629 | 1604 |
| 2012 | 564 | 185 | 49 | 2610 | 579 | 1440 |
| 2013 | 530 | 191 | 39 | 2871 | 506 | 1227 |
| 2014 | 751 | 236 | 1 | 1303 | 391 | 1020 |
| 2015 | 440 | 199 | 0 | 1110 | 317 | 703 |
| 2016 | 305 | 210 | 2 | 547 | 231 | 212 |
| 2017 | 125 | 147 | 0 | 442 | 270 | 216 |
| 2018* | 160 | 267 | 0 | 331 | 61 | 156 |

*Preliminary.

Table 3. Bss.27.4.bc, 7.a, d-h: Sampling of commercial fishery landings of otter, pelagic midwater trawls, lines and nets for length and age by area in the UK (England and Wales). Nsamp = number of landings sampled; Nfish = number of fish.

|  | UK Otter trawl | UK Pelagic/midwater |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age | Length | Landings <br> (t) | Age | Length | Landings <br> (t) |  |  |  |  |
| Year | Nsamp | Nfish | Nsamp | Nfish |  | Nsamp | Nfish | Nsamp | Nfish |  |
| 1985 | 45 | 235 | 15 | 225 | 27 | 3 | 44 | 2 | 43 | 1 |
| 1986 | 18 | 216 | 28 | 2591 | 24 |  |  |  |  | 2 |
| 1987 | 41 | 421 | 54 | 1181 | 41 | 4 | 42 | 1 | 589 | 0.02 |
| 1988 | 23 | 257 | 23 | 1298 | 65 | 2 | 64 | 2 | 1684 | 8 |
| 1989 | 63 | 531 | 44 | 1595 | 80 | 4 | 126 | 4 | 1451 | 7 |
| 1990 | 63 | 883 | 48 | 773 | 67 | 8 | 19 |  |  | 22 |
| 1991 | 92 | 983 | 32 | 731 | 39 | 12 | 125 | 1 | 1490 | 14 |
| 1992 | 69 | 699 | 17 | 398 | 41 | 2 | 50 | 2 | 220 | 8 |
| 1993 | 118 | 1219 | 38 | 836 | 80 | 9 | 39 |  |  | 1 |
| 1994 | 182 | 1927 | 113 | 3925 | 125 |  |  | 1 | 127 | 0.3 |
| 1995 | 28 | 529 | 66 | 1995 | 162 |  |  | 1 | 19 | 4 |
| 1996 | 49 | 660 | 39 | 1041 | 122 | 1 | 41 | 3 | 392 | 87 |
| 1997 | 59 | 1660 | 52 | 2445 | 140 | 1 | 49 |  |  | 71 |
| 1998 | 28 | 676 | 39 | 1442 | 133 | 20 | 95 | 4 | 167 | 85 |
| 1999 | 24 | 379 | 46 | 1216 | 138 | 12 | 382 | 9 | 770 | 220 |
| 2000 | 92 | 759 | 42 | 1814 | 133 | 23 | 847 | 14 | 2463 | 52 |
| 2001 | 45 | 851 | 49 | 2152 | 141 | 3 | 58 | 5 | 691 | 97 |
| 2002 | 54 | 523 | 47 | 1454 | 161 |  |  | 4 | 545 | 110 |
| 2003 | 48 | 512 | 45 | 1418 | 207 | 15 | 459 | 4 | 744 | 127 |
| 2004 | 33 | 361 | 31 | 1295 | 173 | 8 | 161 | 5 | 522 | 131 |
| 2005 | 35 | 498 | 31 | 2432 | 181 | 3 | 149 | 2 | 299 | 68 |
| 2006 | 15 | 252 | 17 | 810 | 160 | 1 | 43 | 1 | 100 | 11 |
| 2007 | 44 | 385 | 21 | 903 | 173 | 1 | 20 | 3 | 355 | 37 |
| 2008 | 37 | 580 | 32 | 2151 | 196 | 6 | 409 | 8 | 1283 | 17 |
| 2009 | 24 | 1184 | 13 | 807 | 175 | 8 | 317 | 6 | 625 | 9 |
| 2010 | 25 | 360 | 28 | 1312 | 150 | 7 | 153 | 3 | 376 | 42 |
| 2011 | 25 | 577 | 49 | 1903 | 137 | 3 | 103 | 4 | 463 | 98 |
| 2012 | 18 | 182 | 41 | 751 | 157 |  |  | 1 | 199 | 49 |
| 2013 | 15 | 289 | 23 | 859 | 125 |  |  |  |  | 39 |
| 2014 | 14 | 164 | 22 | 523 | 104 |  |  |  |  | 1 |
| 2015 | 28 | 377 | 39 | 1277 | 100 | 1 | 4 | 1 | 4 | 1 |
| 2016 | 19 | 256 | 90 | 527 | 52 |  |  |  |  | 2 |
| 2017 | 38 | 510 | 128 | 915 | 51 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 43 | 263 | 43 | 492 | 28 | 1 | 15 | 1 | 33 | 0 |


|  | UK Lines |  |  |  |  | UK Nets |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age |  | Length |  | Landings ( t ) | Age |  | Length |  | Landings ( t ) |
| Year | Nsamp | Nfish | Nsamp | Nfish |  | Nsamp | Nfish | Nsamp | Nfish |  |
| 1985 | 53 | 395 | 19 | 285 | 30 | 34 | 332 | 15 | 181 | 43 |
| 1986 | 60 | 496 | 31 | 894 | 33 | 18 | 251 | 18 | 1132 | 61 |
| 1987 | 92 | 313 | 69 | 557 | 18 | 37 | 528 | 44 | 1321 | 55 |
| 1988 | 66 | 538 | 53 | 1325 | 30 | 37 | 584 | 40 | 1397 | 64 |
| 1989 | 249 | 652 | 26 | 310 | 29 | 49 | 469 | 45 | 1248 | 60 |
| 1990 | 281 | 918 | 22 | 260 | 18 | 24 | 207 | 11 | 456 | 61 |
| 1991 | 346 | 1468 | 53 | 963 | 60 | 57 | 481 | 30 | 583 | 113 |
| 1992 | 418 | 2905 | 111 | 2077 | 23 | 40 | 281 | 28 | 1248 | 64 |
| 1993 | 287 | 1787 | 123 | 1426 | 62 | 127 | 1141 | 94 | 1686 | 66 |
| 1994 | 212 | 1616 | 155 | 3783 | 154 | 146 | 2846 | 157 | 5130 | 229 |
| 1995 | 160 | 1043 | 107 | 1493 | 169 | 95 | 1786 | 150 | 6248 | 262 |
| 1996 | 155 | 1326 | 106 | 1790 | 128 | 85 | 1371 | 113 | 3348 | 186 |
| 1997 | 141 | 1262 | 137 | 2072 | 119 | 73 | 1055 | 106 | 2747 | 195 |
| 1998 | 182 | 1215 | 111 | 2820 | 121 | 88 | 1119 | 82 | 2465 | 108 |
| 1999 | 237 | 1304 | 149 | 3793 | 148 | 127 | 1189 | 74 | 2966 | 137 |
| 2000 | 405 | 1395 | 65 | 1964 | 53 | 119 | 1719 | 104 | 5482 | 103 |
| 2001 | 451 | 2485 | 114 | 2935 | 58 | 140 | 2027 | 92 | 3309 | 122 |
| 2002 | 210 | 1286 | 146 | 3031 | 75 | 220 | 3800 | 206 | 6680 | 201 |
| 2003 | 151 | 1009 | 90 | 3108 | 65 | 171 | 1720 | 224 | 5899 | 146 |
| 2004 | 127 | 906 | 66 | 1980 | 72 | 83 | 974 | 150 | 3567 | 207 |
| 2005 | 87 | 380 | 25 | 921 | 59 | 73 | 768 | 33 | 1126 | 172 |
| 2006 | 54 | 359 | 67 | 989 | 119 | 56 | 598 | 47 | 1197 | 199 |
| 2007 | 94 | 713 | 31 | 1088 | 166 | 90 | 753 | 40 | 1811 | 239 |
| 2008 | 37 | 552 | 28 | 1325 | 163 | 100 | 1444 | 63 | 3361 | 318 |
| 2009 | 49 | 304 | 18 | 915 | 147 | 116 | 1571 | 100 | 3247 | 311 |
| 2010 | 34 | 418 | 40 | 970 | 183 | 63 | 1214 | 66 | 2350 | 302 |
| 2011 | 46 | 1091 | 55 | 2250 | 143 | 34 | 793 | 41 | 1433 | 324 |
| 2012 | 89 | 1295 | 100 | 2215 | 185 | 35 | 909 | 56 | 2809 | 407 |
| 2013 | 41 | 896 | 42 | 1236 | 191 | 42 | 1123 | 49 | 2342 | 405 |
| 2014 | 67 | 1247 | 73 | 1889 | 236 | 60 | 1161 | 71 | 2781 | 647 |
| 2015 | 72 | 1183 | 79 | 3055 | 199 | 48 | 776 | 67 | 3985 | 338 |
| 2016 | 69 | 1151 | 110 | 1236 | 210 | 59 | 1165 | 83 | 1974 | 252 |
| 2017 | 28 | 303 | 171 | 2225 | 158 | 0 | 0 | 41 | 727 | 74 |
| 2018 | 103 | 1478 | 123 | 2166 | 267 | 55 | 694 | 55 | 1763 | 132 |

Table 4. Bss.27.4.bc, 7.a, d-h. Sampling of commercial fishery landings by area in France, giving numbers of fishing trips sampled, number of fish measures, and the total landings ( 2017 based on real sampling, not simulated).

| Year | FR_lines | FR_nets | FR_bottom trawl |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | No. Trips | No.fish | Landings | No. Trips | No.fish | Landings | No. Trips | No.fish | Landings |
| 2000 | 53 | 1613 | 305 | 2 | 72 | 108 | 2 | 196 | 692 |
| 2001 | 101 | 2659 | 375 | 1 | 5 | 110 | 0 | 0 | 713 |
| 2002 | 79 | 2076 | 349 | 0 | 0 | 128 | 4 | 710 | 911 |
| 2003 | 78 | 1732 | 438 | 1 | 4 | 152 | 8 | 998 | 1087 |
| 2004 | 78 | 1748 | 381 | 6 | 84 | 150 | 12 | 887 | 1236 |
| 2005 | 34 | 949 | 439 | 4 | 110 | 148 | 14 | 689 | 1239 |
| 2006 | 73 | 1719 | 554 | 11 | 291 | 140 | 11 | 1240 | 1110 |
| 2007 | 69 | 2235 | 560 | 28 | 641 | 158 | 11 | 588 | 1187 |
| 2008 | 41 | 1280 | 425 | 25 | 496 | 128 | 18 | 1927 | 1145 |
| 2009 | 33 | 1339 | 251 | 25 | 159 | 94 | 93 | 1468 | 1052 |
| 2010 | 10 | 334 | 278 | 49 | 615 | 160 | 64 | 626 | 819 |
| 2011 | 17 | 540 | 359 | 156 | 278 | 129 | 151 | 1955 | 791 |
| 2012 | 10 | 681 | 295 | 60 | 408 | 142 | 87 | 1204 | 824 |
| 2013 | 16 | 309 | 291 | 26 | 512 | 126 | 73 | 2060 | 737 |
| 2014 | 10 | 299 | 285 | 29 | 218 | 163 | 137 | 2139 | 571 |
| 2015 | 16 | 326 | 210 | 35 | 242 | 109 | 76 | 1628 | 642 |
| 2016 | 2 | 84 | 156 | 32 | 293 | 64 | 183 | 1396 | 271 |
| 2017 | 9 | 219 | 166 | 18 | 151 | 35 | 126 | 495 | 33 |
| 2018 | 8 | 357 | 151 | 49 | 312 | 74 | 79 | 274 | 63 |


| Year | FR_pelagic trawl | FR_danish seine | FR_other gears |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | No. Trips | No.fish | Landings | No. Trips | No.fish | Landings | No. Trips | No.fish | Landings |
| 2000 | 2 | 629 | 681 | 0 | 0 | 0 | 0 | 0 | 20 |
| 2001 | 0 | 0 | 659 | 0 | 0 | 0 | 0 | 0 | 27 |
| 2002 | 3 | 680 | 415 | 0 | 0 | 0 | 0 | 0 | 22 |
| 2003 | 4 | 753 | 773 | 0 | 0 | 0 | 0 | 0 | 23 |
| 2004 | 6 | 938 | 820 | 0 | 0 | 0 | 0 | 0 | 17 |
| 2005 | 11 | 1239 | 1319 | 0 | 0 | 0 | 0 | 0 | 17 |
| 2006 | 16 | 2597 | 1420 | 0 | 0 | 0 | 0 | 0 | 35 |
| 2007 | 8 | 1800 | 841 | 0 | 0 | 0 | 0 | 0 | 24 |
| 2008 | 8 | 1065 | 1012 | 0 | 0 | 0 | 0 | 0 | 40 |
| 2009 | 55 | 899 | 1098 | 0 | 0 | 27 | 0 | 0 | 127 |
| 2010 | 28 | 1299 | 1828 | 0 | 0 | 61 | 2 | 2 | 90 |
| 2011 | 30 | 2309 | 1142 | 2 | 6 | 43 | 36 | 292 | 62 |
| 2012 | 9 | 1649 | 1143 | 6 | 370 | 112 | 7 | 154 | 91 |
| 2013 | 10 | 1253 | 1516 | 2 | 28 | 18 | 1 | 1 | 82 |
| 2014 | 23 | 455 | 242 | 12 | 23 | 9 | 1 | 1 | 25 |
| 2015 | 12 | 158 | 107 | 0 | 12 | 26 | 0 | 0 | 16 |
| 2016 | 6 | 48 | 17 | 28 | 78 | 20 | 0 | 0 | 20 |
| 2017 | 0 | 0 | 6 | 14 | 42 | 22 | 0 | 0 | 40 |
| 2018 | 2 | 2 | 1 | 2 | 3 | 9 | 0 | 0 | 16 |

Table 5. Numbers-at-length in French commercial all-gears fishery landings (input to assessment at lengths 14-94 cm).

|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 717 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 26 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 63 | 0 |
| 28 | 0 | 0 | 0 | 3455 | 0 | 0 | 0 | 0 | 0 | 292 | 0 | 0 | 1219 | 0 | 0 | 291 | 0 | 0 | 0 |
| 30 | 0 | 0 | 1015 | 13054 | 14 | 0 | 15689 | 0 | 0 | 473 | 0 | 0 | 0 | 146 | 0 | 346 | 71 | 0 | 0 |
| 32 | 0 | 0 | 0 | 58717 | 13057 | 9903 | 32459 | 181 | 8250 | 2239 | 9811 | 1976 | 1583 | 0 | 3076 | 2678 | 1481 | 0 | 0 |
| 34 | 9931 | 17962 | 12469 | 105655 | 78811 | 29872 | 179130 | 4715 | 28986 | 10714 | 28290 | 13885 | 6518 | 1504 | 3620 | 5102 | 1440 | 137 | 0 |
| 36 | 34932 | 19809 | 38249 | 125326 | 127801 | 97890 | 285704 | 39335 | 229758 | 124925 | 169311 | 57121 | 85760 | 29667 | 33532 | 44175 | 2814 | 2646 | 0 |
| 38 | 85866 | 68920 | 46427 | 180475 | 124051 | 128022 | 217657 | 102714 | 263071 | 211881 | 177571 | 87842 | 172510 | 88507 | 68262 | 75546 | 4340 | 2523 | 91 |
| 40 | 126730 | 76594 | 62503 | 119495 | 227214 | 231750 | 178250 | 146272 | 266408 | 225545 | 182105 | 128838 | 140273 | 149070 | 74871 | 93273 | 7417 | 3572 | 814 |
| 42 | 102836 | 98008 | 82461 | 145456 | 282390 | 266905 | 196868 | 145122 | 237160 | 193030 | 283064 | 187586 | 147895 | 146130 | 82684 | 115713 | 24816 | 9257 | 2444 |
| 44 | 80478 | 109595 | 91064 | 104545 | 243107 | 344681 | 289998 | 164011 | 270810 | 222613 | 251956 | 201447 | 162333 | 123170 | 51365 | 122460 | 20422 | 14861 | 2954 |
| 46 | 93344 | 106857 | 86723 | 130023 | 188494 | 270532 | 285451 | 130859 | 228996 | 238849 | 230227 | 199487 | 180752 | 140677 | 61292 | 95208 | 22427 | 9603 | 4379 |
| 48 | 80934 | 77694 | 62163 | 115806 | 126685 | 239265 | 263272 | 100043 | 142650 | 155222 | 188149 | 194697 | 158490 | 127136 | 39844 | 59668 | 20653 | 7367 | 2606 |
| 50 | 55399 | 57055 | 55905 | 91915 | 72581 | 169478 | 200874 | 99210 | 112385 | 159658 | 186310 | 145447 | 130759 | 116842 | 38109 | 51436 | 15619 | 6801 | 3549 |
| 52 | 52948 | 51658 | 46180 | 93878 | 82331 | 115269 | 119836 | 75929 | 74336 | 114530 | 109212 | 124239 | 107214 | 99156 | 29929 | 37860 | 10415 | 4599 | 2861 |


|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 54 | 42094 | 36737 | 35998 | 48742 | 50633 | 62106 | 99509 | 74405 | 66260 | 84649 | 120550 | 92526 | 90638 | 103818 | 39911 | 21406 | 16034 | 3586 | 2702 |
| 56 | 26460 | 35839 | 26001 | 60839 | 60284 | 67741 | 99674 | 55147 | 48853 | 96257 | 71590 | 72471 | 78934 | 89197 | 32298 | 20681 | 9753 | 1012 | 2538 |
| 58 | 27357 | 22762 | 19019 | 31614 | 31334 | 61132 | 54522 | 46087 | 39689 | 51578 | 62211 | 46869 | 54869 | 59004 | 30016 | 13591 | 12328 | 2519 | 3581 |
| 60 | 23581 | 25834 | 14210 | 33688 | 19126 | 43591 | 45908 | 28056 | 29840 | 36547 | 31544 | 31690 | 35387 | 65851 | 21467 | 11946 | 7678 | 913 | 2008 |
| 62 | 14295 | 18773 | 11129 | 30691 | 23996 | 35774 | 23763 | 23057 | 28335 | 57472 | 19076 | 19998 | 33085 | 64579 | 16797 | 11776 | 7506 | 1120 | 1669 |
| 64 | 18044 | 13532 | 16771 | 18823 | 14799 | 25788 | 20607 | 18091 | 14420 | 24016 | 62005 | 17624 | 17714 | 53482 | 16261 | 9356 | 4348 | 1369 | 1641 |
| 66 | 10773 | 11068 | 11011 | 13230 | 10650 | 12456 | 14969 | 8715 | 12694 | 21415 | 26388 | 14720 | 15170 | 37744 | 8387 | 6653 | 2634 | 510 | 778 |
| 68 | 9903 | 9120 | 5447 | 7960 | 8569 | 13360 | 13976 | 8793 | 9039 | 27466 | 9340 | 7906 | 9374 | 23884 | 5579 | 2485 | 4465 | 315 | 463 |
| 70 | 5709 | 11771 | 4795 | 5374 | 4880 | 8908 | 9653 | 4835 | 6821 | 20198 | 8541 | 6114 | 8114 | 32512 | 8995 | 1163 | 1353 | 345 | 255 |
| 72 | 5721 | 5733 | 4559 | 5617 | 2974 | 8053 | 4521 | 2707 | 4714 | 12083 | 29128 | 2082 | 4147 | 14996 | 3027 | 660 | 956 | 408 | 47 |
| 74 | 2345 | 5345 | 1825 | 3275 | 2675 | 9811 | 3424 | 1962 | 1623 | 7551 | 1884 | 1163 | 2313 | 9001 | 642 | 628 | 219 | 652 | 0 |
| 76 | 2595 | 2782 | 1260 | 1356 | 2567 | 5020 | 2883 | 1010 | 1257 | 979 | 2114 | 1096 | 1540 | 2640 | 773 | 431 | 0 | 92 | 0 |
| 78 | 2102 | 1691 | 357 | 297 | 548 | 2378 | 731 | 399 | 534 | 1765 | 182 | 476 | 1134 | 2073 | 0 | 9 | 127 | 718 | 0 |
| 80 | 888 | 583 | 155 | 783 | 425 | 1365 | 201 | 158 | 261 | 264 | 5525 | 148 | 282 | 176 | 198 | 16 | 0 | 0 | 0 |
| 82 | 1021 | 296 | 109 | 112 | 149 | 107 | 261 | 37 | 8 | 1004 | 6097 | 104 | 451 | 1566 | 0 | 278 | 0 | 92 | 0 |
| 84 | 548 | 204 | 0 | 148 | 295 | 0 | 30 | 59 | 0 | 0 | 863 | 0 | 29 | 0 | 0 | 0 | 0 | 0 | 0 |
| 86 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 27 | 1115 | 0 | 0 | 301 | 0 | 0 |
| 88 | 0 | 61 | 0 | 0 | 149 | 0 | 0 | 0 | 0 | 0 | 1207 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 6. Numbers-at-age in French commercial fishery landings, 2000-2018, all gears combined.

|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age2 | 0 | 0 | 0 | 2611 | 3 | 0 | 3138 | 0 | 1208 | 315 | 717 | 0 | 0 | 0 | 0 | 47 | 24 | 21 | 0 |
| age3 | 0 | 2651 | 8114 | 10800 | 4 | 24195 | 74600 | 5307 | 79917 | 23355 | 1962 | 0 | 406 | 60 | 603 | 1394 | 565 | 105 | 0 |
| age4 | 9440 | 55640 | 73892 | 364427 | 80483 | 77794 | 131099 | 73224 | 175402 | 119979 | 39409 | 6087 | 14357 | 569 | 6846 | 20917 | 3419 | 1154 | 356 |
| age5 | 222655 | 47734 | 125531 | 241694 | 627951 | 253455 | 564668 | 135809 | 545960 | 282754 | 221063 | 172404 | 65157 | 52216 | 11735 | 116939 | 23364 | 5712 | 1325 |
| age6 | 273687 | 298773 | 90294 | 318445 | 438799 | 735235 | 361515 | 460583 | 401231 | 473020 | 515711 | 252236 | 262593 | 96064 | 123435 | 139446 | 25335 | 16085 | 4079 |
| age7 | 139562 | 211740 | 236147 | 96562 | 297961 | 352182 | 841651 | 124606 | 456312 | 238022 | 411737 | 312186 | 346334 | 609903 | 149938 | 125305 | 22790 | 11655 | 6239 |
| age8 | 79413 | 90962 | 86108 | 254050 | 65297 | 443765 | 146484 | 139879 | 143871 | 408951 | 437222 | 303804 | 308183 | 377156 | 133129 | 191220 | 29076 | 11830 | 5199 |
| age9 | 47258 | 44742 | 31151 | 114829 | 131612 | 39104 | 253945 | 79978 | 147881 | 100487 | 200328 | 314164 | 264012 | 367869 | 143241 | 88543 | 38383 | 7528 | 4401 |
| age10 | 43924 | 21074 | 23025 | 57883 | 77533 | 161572 | 13655 | 69214 | 40719 | 200417 | 172430 | 125800 | 214803 | 481247 | 39242 | 67528 | 26822 | 3485 | 4393 |
| age11 | 49293 | 39908 | 17823 | 26223 | 25416 | 69617 | 132370 | 33191 | 57341 | 73570 | 109342 | 89188 | 83939 | 245982 | 39476 | 24658 | 18455 | 3636 | 3087 |
| age12 | 20207 | 36007 | 14760 | 19879 | 14848 | 26314 | 84910 | 65868 | 17882 | 37114 | 75421 | 34465 | 50701 | 158757 | 12679 | 17551 | 4964 | 3571 | 2340 |
| age13 | 10767 | 17787 | 15912 | 14232 | 14254 | 17996 | 22068 | 68599 | 35092 | 32657 | 46461 | 28352 | 24784 | 43008 | 7347 | 5046 | 3114 | 1819 | 1358 |
| age14 | 4925 | 4394 | 9752 | 18088 | 13528 | 19238 | 6648 | 11131 | 12669 | 55506 | 21880 | 12942 | 8470 | 21825 | 3067 | 5387 | 1866 | 1544 | 638 |
| age15 | 4927 | 6838 | 3743 | 6600 | 7628 | 17974 | 6999 | 9034 | 5518 | 33537 | 4806 | 5585 | 3191 | 14812 | 198 | 431 | 381 | 0 | 275 |
| age16 | 10901 | 8034 | 1553 | 4028 | 5270 | 22718 | 16069 | 5486 | 6091 | 23529 | 16480 | 337 | 1583 | 11520 | 0 | 428 | 429 | 0 | 329 |

Table 7. Bss.27.4.bc, 7.a, d-h. Numbers-at-age in UK(England and Wales) mixed bottom otter trawl, nets.

|  | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 65 | 11844 | 30828 | 6121 | 9692 | 1240 | 3914 | 9713 | 2454 | 2581 | 1320 | 343 | 841 | 286 | 892 |
| 1986 | 0 | 15673 | 20303 | 18759 | 3453 | 7662 | 704 | 3197 | 10503 | 1833 | 1403 | 2889 | 1222 | 1688 | 3595 |
| 1987 | 0 | 439 | 30263 | 58458 | 13753 | 2095 | 2437 | 656 | 726 | 5731 | 2565 | 1889 | 761 | 817 | 2796 |
| 1988 | 0 | 1930 | 20862 | 54472 | 41710 | 12803 | 1721 | 2315 | 780 | 451 | 5503 | 2024 | 1312 | 801 | 2589 |
| 1989 | 33394 | 5411 | 1223 | 7659 | 43911 | 26891 | 9002 | 3076 | 2901 | 1878 | 2896 | 8914 | 1499 | 1286 | 3436 |
| 1990 | 0 | 3035 | 2503 | 3770 | 16047 | 31459 | 21020 | 5042 | 2186 | 1463 | 846 | 1100 | 4837 | 353 | 2703 |
| 1991 | 1533 | 6933 | 36938 | 2381 | 1283 | 6576 | 18064 | 16248 | 7033 | 589 | 2617 | 2321 | 480 | 6659 | 3674 |
| 1992 | 0 | 15982 | 55550 | 33557 | 1183 | 796 | 1956 | 4750 | 4762 | 1230 | 451 | 433 | 139 | 497 | 3202 |
| 1993 | 0 | 657 | 81429 | 65981 | 21858 | 1351 | 627 | 1796 | 4803 | 3920 | 1500 | 710 | 735 | 475 | 2347 |
| 1994 | 2 | 1328 | 30970 | 369416 | 41472 | 16079 | 1130 | 294 | 2282 | 5842 | 4387 | 1596 | 650 | 646 | 3717 |
| 1995 | 0 | 5599 | 37064 | 81529 | 334815 | 17932 | 6931 | 702 | 415 | 1046 | 3440 | 3215 | 1846 | 2699 | 2680 |
| 1996 | 191 | 11473 | 43831 | 31632 | 64618 | 173733 | 8235 | 3622 | 216 | 315 | 454 | 1881 | 1688 | 534 | 1784 |
| 1997 | 0 | 2490 | 8501 | 64000 | 45238 | 39229 | 145407 | 8105 | 4456 | 632 | 640 | 294 | 2689 | 1712 | 2235 |
| 1998 | 0 | 1103 | 44997 | 49461 | 69489 | 25366 | 15136 | 41057 | 2671 | 860 | 96 | 96 | 385 | 623 | 811 |
| 1999 | 241 | 82 | 80414 | 146338 | 43841 | 28582 | 9612 | 6192 | 18072 | 1112 | 729 | 40 | 270 | 97 | 830 |
| 2000 | 0 | 9528 | 2584 | 151515 | 72747 | 11772 | 11046 | 4992 | 4636 | 8323 | 818 | 184 | 14 | 55 | 643 |
| 2001 | 614 | 11085 | 92408 | 29064 | 105169 | 25329 | 7388 | 8742 | 5811 | 8136 | 7522 | 804 | 768 | 69 | 759 |
| 2002 | 338 | 11495 | 43605 | 240476 | 16779 | 67647 | 16021 | 7450 | 8022 | 2682 | 3842 | 10166 | 645 | 193 | 568 |


|  | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 |
| :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 8. Bss.27.4.bc, 7.a, d-h: Numbers-at-age in UK(England and Wales) Lines.

|  | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 0 | 577 | 8939 | 3343 | 933 | 2354 | 358 | 758 | 5428 | 960 | 871 | 953 | 573 | 645 | 1307 |
| 1987 | 0 | 108 | 1052 | 3719 | 2132 | 581 | 477 | 432 | 523 | 1578 | 845 | 211 | 167 | 179 | 1187 |
| 1988 | 0 | 33 | 1751 | 13389 | 5067 | 2398 | 551 | 1014 | 209 | 456 | 1863 | 895 | 715 | 523 | 977 |
| 1989 | 22 | 0 | 538 | 8171 | 36046 | 1842 | 371 | 104 | 208 | 58 | 215 | 1040 | 115 | 87 | 334 |
| 1990 | 0 | 305 | 82 | 185 | 1284 | 3456 | 2407 | 897 | 357 | 369 | 193 | 242 | 1261 | 81 | 828 |
| 1991 | 0 | 131 | 8420 | 471 | 177 | 792 | 4927 | 4024 | 1842 | 89 | 1229 | 1685 | 367 | 4831 | 2887 |
| 1992 | 0 | 1195 | 5473 | 5267 | 294 | 269 | 518 | 1193 | 1633 | 563 | 130 | 195 | 169 | 143 | 1411 |
| 1993 | 16 | 526 | 11652 | 11776 | 7569 | 590 | 289 | 931 | 3941 | 3344 | 1367 | 663 | 703 | 643 | 3789 |
| 1994 | 0 | 71 | 4059 | 119784 | 18540 | 9393 | 943 | 173 | 1754 | 5414 | 5570 | 1205 | 639 | 274 | 2790 |
| 1995 | 0 | 486 | 6943 | 21979 | 97509 | 7380 | 5313 | 480 | 699 | 831 | 5684 | 3696 | 1936 | 840 | 4733 |
| 1996 | 0 | 210 | 8804 | 12487 | 15338 | 57127 | 4566 | 4979 | 127 | 510 | 364 | 2521 | 1573 | 1300 | 2346 |
| 1997 | 59 | 454 | 3102 | 15613 | 11415 | 8287 | 50819 | 2853 | 1635 | 557 | 354 | 243 | 2195 | 1065 | 1570 |
| 1998 | 0 | 3676 | 8366 | 10920 | 22630 | 10485 | 6452 | 28231 | 2949 | 1091 | 138 | 196 | 793 | 1381 | 1254 |
| 1999 | 479 | 255 | 25158 | 37306 | 13589 | 13697 | 5288 | 5001 | 20522 | 1669 | 2038 | 247 | 777 | 315 | 3314 |
| 2000 | 0 | 421 | 294 | 19380 | 12402 | 2696 | 3285 | 1476 | 1248 | 4697 | 330 | 258 | 16 | 88 | 559 |
| 2001 | 54 | 471 | 7385 | 1392 | 17864 | 7702 | 2027 | 3239 | 1685 | 1761 | 3774 | 440 | 301 | 27 | 420 |
| 2002 | 30 | 729 | 2609 | 14173 | 2686 | 17358 | 7757 | 2621 | 5179 | 1463 | 1766 | 3687 | 322 | 101 | 180 |
| 2003 | 0 | 80 | 7166 | 7917 | 25014 | 2167 | 10164 | 3262 | 1473 | 982 | 796 | 681 | 1704 | 186 | 166 |


|  | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 0 | 279 | 1697 | 13884 | 8601 | 17310 | 2398 | 6365 | 3626 | 1181 | 1189 | 1172 | 406 | 2243 | 143 |
| 2005 | 0 | 621 | 2669 | 5059 | 14699 | 5529 | 6985 | 589 | 5697 | 1845 | 236 | 1307 | 33 | 189 | 606 |
| 2006 | 0 | 44 | 16121 | 35990 | 13714 | 22306 | 5794 | 12717 | 1644 | 3135 | 1258 | 305 | 358 | 1016 | 734 |
| 2007 | 0 | 22 | 6611 | 31578 | 28396 | 14511 | 17834 | 8499 | 10951 | 5163 | 3121 | 5119 | 85 | 344 | 485 |
| 2008 | 0 | 199 | 5010 | 27319 | 42071 | 21561 | 12265 | 12566 | 5458 | 4960 | 1372 | 1032 | 3431 | 198 | 992 |
| 2009 | 0 | 315 | 8415 | 19843 | 33661 | 25695 | 12017 | 9320 | 5021 | 5371 | 4748 | 811 | 1075 | 0 | 0 |
| 2010 | 0 | 814 | 7029 | 45515 | 54766 | 39716 | 15835 | 5147 | 2395 | 2910 | 706 | 522 | 359 | 81 | 277 |
| 2011 | 0 | 8 | 5209 | 11538 | 24667 | 19293 | 16668 | 13032 | 4947 | 6066 | 2695 | 1941 | 2187 | 522 | 657 |
| 2012 | 0 | 91 | 1695 | 18362 | 28593 | 23507 | 22946 | 17909 | 10199 | 7725 | 2994 | 2672 | 2158 | 596 | 820 |
| 2013 | 0 | 0 | 1187 | 6979 | 35135 | 32251 | 18057 | 14762 | 10333 | 10543 | 6106 | 3730 | 2886 | 1957 | 1938 |
| 2014 | 0 | 980 | 4985 | 26081 | 20743 | 39548 | 28357 | 15323 | 12440 | 12413 | 8018 | 4889 | 1976 | 1673 | 1322 |
| 2015 | 0 | 6 | 1834 | 5941 | 23369 | 22221 | 31442 | 19014 | 10344 | 8210 | 7036 | 2504 | 3136 | 744 | 798 |
| 2016 | 0 | 0 | 742 | 7020 | 11858 | 20142 | 15479 | 25838 | 13362 | 7406 | 5904 | 4674 | 2548 | 3894 | 2567 |
| 2017 | 0 | 0 | 1734 | 4007 | 5766 | 2324 | 2362 | 1036 | 4159 | 993 | 356 | 469 | 202 | 475 | 330 |
| 2018 | 0 | 454 | 6992 | 23652 | 41538 | 31173 | 17352 | 16753 | 11214 | 14117 | 9044 | 4650 | 3791 | 2220 | 3945 |

Table 9. Bss.27.4.bc, 7.a, d-h: Numbers-at-age in UK (England and Wales) midwater pair trawl fleet (no samples for 1997, 2013, 2014, 2015, 2016, 2017).

|  | Age3 | Age4 | Age5 | Age 6 | Age 7 | Age 8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 0 | 289 | 796 | 3892 | 71666 | 5583 | 1648 | 21 | 334 | 154 | 622 | 485 | 199 | 559 |
| 1998 | 0 | 245 | 5979 | 11845 | 8553 | 8135 | 25138 | 2517 | 345 | 93 | 53 | 119 | 893 | 569 |
| 1999 | 0 | 2983 | 18409 | 15106 | 27147 | 13818 | 18060 | 43097 | 4389 | 1686 | 324 | 387 | 308 | 2689 |
| 2000 | 15 | 60 | 2476 | 7587 | 3270 | 4497 | 1459 | 2830 | 7077 | 634 | 174 | 39 | 96 | 420 |
| 2001 | 0 | 179 | 899 | 19777 | 20290 | 7042 | 5268 | 3124 | 2845 | 9666 | 857 | 636 | 123 | 261 |
| 2002 | 3 | 37 | 2380 | 1578 | 24087 | 9693 | 6297 | 5978 | 450 | 5664 | 9215 | 0 | 0 | 530 |
| 2003 | 0 | 2689 | 10619 | 39257 | 7971 | 40551 | 10293 | 3162 | 3254 | 618 | 169 | 4043 | 77 | 281 |
| 2004 | 7 | 1254 | 12502 | 14372 | 48109 | 3199 | 20694 | 8010 | 353 | 1797 | 1141 | 91 | 968 | 18 |
| 2005 | 0 | 114 | 2103 | 15321 | 14397 | 17408 | 1907 | 5182 | 0 | 1831 | 99 | 0 | 40 | 599 |
| 2006 | 0 | 227 | 567 | 608 | 4076 | 1423 | 3085 | 254 | 176 | 111 | 0 | 0 | 0 | 53 |
| 2007 | 0 | 385 | 2517 | 7038 | 5387 | 6833 | 2795 | 1900 | 631 | 807 | 12 | 37 | 19 | 121 |
| 2008 | 45 | 445 | 1540 | 3279 | 1787 | 1412 | 1557 | 755 | 960 | 30 | 183 | 490 | 0 | 40 |
| 2009 | 0 | 90 | 635 | 2175 | 2596 | 843 | 784 | 168 | 298 | 173 | 11 | 169 | 0 | 0 |
| 2010 | 9 | 36 | 1741 | 5546 | 8261 | 6678 | 4755 | 403 | 3786 | 152 | 294 | 313 | 551 | 50 |
| 2011 | 0 | 255 | 4397 | 10231 | 13640 | 15909 | 13642 | 4424 | 4233 | 2773 | 1688 | 1003 | 264 | 423 |
| 2012 | 0 | 391 | 4461 | 10776 | 10016 | 8757 | 5789 | 2741 | 1134 | 290 | 433 | 143 | 127 | 226 |
| 2015 | 0 | 7 | 23 | 85 | 103 | 137 | 30 | 6 | 3 | 0 | 0 | 0 | 0 | 0 |

## Table 10. Bss.27.4.bc, 7.a, d-h: Numbers of trips sampled for discards by Cefas (UK): 2002-2015, by gear group and area.

| Division (a) bottom otter trawls | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 16 | 34 | 56 | 37 | 41 | 85 | 58 | 49 | 46 | 42 | 54 | 30 | 53 | 45 | 12 | 0 | 1 |
| 7.afg | 8 | 15 | 23 | 8 | 11 | 43 | 50 | 28 | 22 | 22 | 22 | 12 | 14 | 16 | 2 | 0 | 0 |
| 7.d | 1 | 2 | 4 | 3 | 1 | 2 | 1 | 6 | 7 | 9 | 4 | 5 | 7 | 3 | 13 | 1 | 1 |
| 7.eh | 9 | 24 | 37 | 31 | 49 | 90 | 87 | 38 | 29 | 32 | 29 | 45 | 73 | 68 | 29 | 0 | 10 |
| total | 34 | 75 | 120 | 79 | 102 | 220 | 196 | 121 | 104 | 105 | 109 | 92 | 147 | 132 | 56 | 1 | 12 |
| (b) Fixed/driftnets |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Division | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| 4 | 0 | 0 | 2 | 1 | 11 | 31 | 15 | 20 | 15 | 11 | 13 | 18 | 10 | 7 | 0 | 0 | 0 |
| 7.afg | 3 | 7 | 5 | 3 | 7 | 8 | 9 | 10 | 7 | 16 | 22 | 16 | 25 | 12 | 3 | 0 | 0 |
| 7.d | 0 | 0 | 1 | 0 | 0 | 17 | 6 | 4 | 1 | 7 | 10 | 42 | 25 | 17 | 10 | 0 | 0 |
| 7.eh | 1 | 5 | 9 | 2 | 3 | 16 | 10 | 14 | 19 | 17 | 25 | 24 | 24 | 15 | 0 | 0 | 0 |
| total | 4 | 12 | 17 | 6 | 21 | 72 | 40 | 48 | 42 | 51 | 70 | 100 | 84 | 51 | 13 | 0 | 0 |
| (c) Lines |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Division | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| 4 | 0 | 1 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
| 7.afg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 7.d | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 33 | 2 | 0 |
| 7.eh | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 8 | 5 | 4 | 0 | 0 |


| total | 0 | 1 | 1 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 2 | 1 | 10 | 6 | 37 | 2 | 0 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

(d) Midwater trawls

| Division | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.afg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.d | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.eh | 0 | 1 | 1 | 1 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 |
| total | 1 | 1 | 1 | 3 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 2 |  | 0 | 0 | 0 | 0 |

(e) Other gears

| Division | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 8 | 5 | 10 | 1 | 2 | 1 | 1 | 7 | 6 | 8 | 4 | 10 |  | 0 | 6 | 0 | 0 |
| 7.afg | 4 | 11 | 8 | 4 | 9 | 1 | 2 | 3 | 3 | 1 | 4 | 8 |  | 0 | 5 | 0 | 0 |
| 7.d | 0 | 1 | 5 | 2 | 3 | 1 | 1 | 2 | 4 | 1 | 2 | 3 | 1 | 2 | 0 | 0 | 0 |
| 7.eh | 10 | 17 | 27 | 16 | 24 | 32 | 18 | 13 | 17 | 27 | 22 | 21 | 14 | 15 | 1 | 0 | 0 |
| total | 22 | 34 | 50 | 23 | 38 | 35 | 22 | 25 | 30 | 37 | 32 | 42 | 15 | 17 | 12 | 0 | 0 |

Table 11. Bss.27.4.bc, 7.a, d-h: Estimated annual numbers and weight of seabass retained and discarded by UK using fixed or driftnets, otter trawl, beam trawl and lines fleets in areas 4, 7.d, 7.eh and 7.afg, based on at-sea sampling, and raised from landings in sampled strata to landings in all strata. Numbers of sampled trips are shown.

|  | Otter trawl | Nets | Beam trawl | Lines | Total ОTB, nets, lines and BTS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | discards | retained | rate <br> (\%) | No. trips sampled | discards | retained | rate \% | No. trips sampled | discards | retained | rate <br> \% | No. trips sampled | discards | retained | rate <br> (\%) | No. trips sampled | discards | retained | rate\% |
| 2002 | 17 | 161 | 9 | 34 | 0 | 201 | 0 | 4 | 0.2 | 24 | 0.7 | - | - | - | - | - | 17 | 386 | 4 |
| 2003 | 16 | 207 | 7 | 75 | 0 | 146 | 0 | 12 | 1.9 | 21 | 8.1 | - | - | - | - | - | 18 | 374 | 5 |
| 2004 | 59 | 173 | 25 | 120 | 0 | 207 | 0 | 17 | 0.3 | 24 | 1.3 | - | - | - | - | - | 59 | 404 | 13 |
| 2005 | 6 | 181 | 3 | 79 | 90 | 172 | 34 | 6 | 2.4 | 15 | 13.7 | - | - | - | - | - | 99 | 368 | 21 |
| 2006 | 34 | 160 | 17 | 102 | 19 | 199 | 9 | 21 | 0.4 | 14 | 2.5 | - | - | - | - | - | 53 | 373 | 12 |
| 2007 | 49 | 173 | 22 | 220 | 1 | 239 | 0.4 | 72 | 0.0 | 19 | 0.0 | - | - | - | - | - | 50 | 432 | 10 |
| 2008 | 5 | 196 | 3 | 196 | 3 | 318 | 0.9 | 40 | 1.2 | 21 | 5.6 | - | - | - | - | - | 9 | 535 | 2 |
| 2009 | 85 | 175 | 33 | 121 | 0 | 311 | 0.1 | 48 | 0.2 | 10 | 1.5 | - | - | - | - | - | 86 | 495 | 15 |
| 2010 | 49 | 150 | 25 | 104 | 1 | 302 | 0.3 | 42 | 1.2 | 6 | 17.1 | - | - | - | - | - | 51 | 458 | 10 |
| 2011 | 8 | 137 | 6 | 105 | 14 | 324 | 4.2 | 51 | 0.0 | 5 | 0.0 | - | - | - | - | - | 22 | 467 | 5 |
| 2012 | 27 | 157 | 15 | 109 | 2 | 407 | 0.5 | 70 | 0.0 | 5 | 0.0 | - | - | - | - | - | 29 | 569 | 5 |
| 2013 | 4 | 125 | 3 | 92 | 2 | 405 | 0.4 | 100 | 1.1 | 4 | 20.1 | - | - | - | - | - | 6 | 534 | 1 |
| 2014 | 1 | 104 | 1 | 147 | 6 | 647 | 0.9 | 84 | 0.0 | 8 | 0.0 | - | - | - | - | - | 7 | 758 | 1 |
| 2015 | 6 | 77 | 7 | 132 | 1 | 340 | 0.4 | 51 | 0.0 | 8 | 0.0 | - | - | - | - | - | 7 | 425 | 2 |
| 2016 | 35 | 52 | 40 | 56 | 8 | 252 | 3 | 13 | 0.1 | 23 | 0.0 |  | 8.4 | 210.0 | 4.0 | 37.0 | 52 | 537 | 9 |
| 2017* | 0 | 35 | 1 | 1 | - | 74 | - | 0 | - | 16 | - | 0 | 11 | 147 | 7 | 2 |  | 272 | - |
| 2018* | 11 | 13 | 46 | 5 | - | 132 | - | 0 | 15 | 13 | 54 | 7 | - | 267 | - | 0 | 26 | 425 | 6 |

*Not used in assessment (lack of information. High probability of underestimation considering management measures).

Table 12. Bss.27.4.bc, 7.a, d-h: Number of fishing trips sampled for retained and discarded weight of seabass on French vessels using different gear types: 2009-2018. (Data are clearly underestimated from 2015 and are not used in assessment).

| pelagic trawl FR | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0 | 773 | 0.00\% | NA |  |  |
| 2004 | 0 | 820 | 0.00\% | NA |  |  |
| 2005 | 0 | 1319 | 0.00\% | NA |  |  |
| 2006 | 0 | 1420 | 0.00\% | NA |  |  |
| 2007 | 0 | 841 | 0.00\% | NA | 12 | 2 |
| 2008 | 2 | 1012 | 0.20\% | 3.93 | 21 | 4 |
| 2009 | 21.2 | 1098 | 1.89\% | 0.05 |  |  |
| 2010 | 7.4 | 1828 | 0.40\% | 0.71 | 35 | 106 |
| 2011 | 7.2 | 1142 | 0.63\% | 0.12 | 9 | 46 |
| 2012 | 0.9 | 1143 | 0.08\% | 2.38 | 7 | 29 |
| 2013 | 0.3 | 1516 | 0.02\% | 2 |  |  |
| 2014 | 0 | 242 | 0.00\% | NA |  |  |
| 2015 | 11.7 | 107 | 9.86\% | 0.03 | 32 | 5 |
| 2016 | 0.5 | 17.43081 | 2.79\% | NA | 19 | 2 |
| 2017 |  | 6 |  | NA | 0 | 0 |
| 2018 | 0.2 | 1 | 17\% |  | 28 | 1 |


| bottom trawlFR | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 73.8 | 1087 | 6.36\% | 0.35 | 18 | 26 |
| 2004 |  | 1236 | NA | NA | 24 | 3 |
| 2005 | 43.9 | 1239 | 3.42\% | 0.9 |  |  |
| 2006 | 42.9 | 1110 | 3.72\% | 1.07 | 24 | 36 |
| 2007 | 9.6 | 1187 | 0.80\% | 0.73 |  |  |
| 2008 | 40.7 | 1145 | 3.43\% | 0.94 | 57 | 63 |
| 2009 |  | 1052 | NA | NA | 143 | 102 |
| 2010 | 76.6 | 819 | 8.55\% | 0.32 | 137 | 5 |
| 2011 | 27.2 | 791 | 3.32\% | 0.46 | 122 | 57 |
| 2012 | 24.5 | 824 | 2.89\% | 0.23 | 151 | 118 |
| 2013 | 26.3 | 737 | 3.45\% | 0.37 | 139 | 145 |
| 2014 |  | 571 | NA | NA | 133 | 29 |
| 2015 | 35.4 | 642 | 5.23\% | 0.49 | 189 | 356 |
| 2016 | 126.9 | 271 | 31.86\% | NA | 512 | 90 |
| 2017 | 156 | 178 | 47\% | NA | 61 | 141 |
| 2018 | 32 | 72 | $31 \%$ |  | 217 | 71 |


| netsFR | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 31.7 | 152 | 17.26\% | 1.2 |  |  |
| 2004 | 77.6 | 150 | 34.09\% | 0.1 |  |  |
| 2005 | 0 | 148 | 0.00\% | NA |  |  |
| 2006 | 125.5 | 140 | 47.27\% | 0.34 |  |  |
| 2007 | 2.2 | 158 | 1.37\% | 0.61 | 32 | 2 |
| 2008 | 0.5 | 128 | 0.39\% | 0.79 |  |  |
| 2009 | 6.4 | 94 | 6.37\% | 0.41 | 196 | 3 |
| 2010 | 6.1 | 160 | 3.67\% | 0.29 | 108 | 5 |
| 2011 | 9 | 129 | 6.52\% | 0.35 |  |  |
| 2012 | 11.8 | 142 | 7.67\% | 0.55 | 269 | 9 |
| 2013 | 21.6 | 126 | 14.63\% | 0.18 | 173 | 2 |
| 2014 | 21.7 | 163 | 11.75\% | 0.11 | 118 | 3 |
| 2015 | 14.7 | 109 | 11.88\% | 0.2 | 217 | 8 |
| 2016 | 19.4 | 64 | 23.25\% | NA | 258 | 209 |
| 2017 | 0.7 | 34 | 2\% | NA | 0 | 0 |
| 2018 | 2 | 74 | 3\% |  | 101 | 17 |


| linesFR | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0 | 438 | 0.00\% | NA |  |  |
| 2004 | 0 | 381 | 0.00\% | NA |  |  |
| 2005 | 0 | 439 | 0.00\% | NA |  |  |
| 2006 | 0 | 554 | 0.00\% | NA |  |  |
| 2007 | 0 | 560 | 0.00\% | NA |  |  |
| 2008 | 100.3 | 425 | 19.09\% | 0.35 |  |  |
| 2009 | 5.6 | 251 | 2.18\% | 0.71 | 17 | 21 |
| 2010 | 3.9 | 278 | 1.38\% | 1.24 |  |  |
| 2011 | 13.1 | 359 | 3.52\% | 0.35 |  |  |
| 2012 | 15.8 | 295 | 5.08\% | 0.26 |  |  |
| 2013 | 14.2 | 291 | 4.65\% | 0.45 |  |  |
| 2014 | 15.8 | 285 | 5.25\% | 0.4 |  |  |
| 2015 | 7.4 | 210 | 3.40\% | 0.32 | 28 | 21 |
| 2016 |  | 156 |  | NA |  |  |
| 2017 |  | 166 |  | NA | 0 | 0 |
| 2018 |  | 151 |  |  | 0 | 0 |


| OtherFR | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0 | 23 | 0.00\% | NA |  |  |
| 2004 | 6.6 | 17 | 27.97\% | NA |  |  |
| 2005 | 0 | 17 | 0.00\% | NA |  |  |
| 2006 | 0 | 35 | 0.00\% | NA |  |  |
| 2007 | 0 | 24 | 0.00\% | NA |  |  |
| 2008 | 0 | 40 | NA | NA |  |  |
| 2009 | 0 | 127 | NA | NA |  |  |
| 2010 | 0 | 90 | 0.00\% | NA |  |  |
| 2011 | 44.8 | 62 | 41.95\% | 5.97 |  |  |
| 2012 | 1.1 | 91 | 1.19\% | 0.25 | 6 | 9 |
| 2013 | 0 | 82 | 0.00\% | NA |  |  |
| 2014 | 0 | 25 | 0.00\% | NA | 130 | 96 |
| 2015 | 11 | 11 | 50.00\% | 0.58 |  |  |
| 2016 | 5.9 | 19.82406 | 22.94\% | NA | 64 | 9 |
| 2017 | 5 | 58 | 8\% | NA | 0 | 0 |
| 2018 |  | 15 |  |  | 0 | 0 |


| FR_ALL | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 105.5 | 2473 | 4\% |  | 18 | 26 |
| 2004 | 84.2 | 2604 | 3\% |  | 24 | 3 |
| 2005 | 43.9 | 3162 | 1\% |  | 0 | 0 |
| 2006 | 168.4 | 3259 | 5\% |  | 24 | 36 |
| 2007 | 11.8 | 2770 | 0\% |  | 44 | 4 |
| 2008 | 143.5 | 2750 | 5\% |  | 78 | 67 |
| 2009 | 33.2 | 2622 | 1\% |  | 356 | 126 |
| 2010 | 94 | 3175 | 3\% |  | 280 | 116 |
| 2011 | 101.3 | 2483 | 4\% | 7.25 | 131 | 103 |
| 2012 | 54.1 | 2495 | 2\% | 3.67 | 433 | 165 |
| 2013 | 62.4 | 2752 | 2\% |  | 312 | 147 |
| 2014 | 37.5 | 1286 | 3\% |  | 381 | 128 |
| 2015 | 80.2 | 1079 | 7\% | 1.62 | 466 | 390 |
| 2016 | 152.7 | 529 | 22\% |  | 853 | 310 |
| 2017 | 161.7 | 442 | 27\% |  | 61 | 141 |
| 2018 | 34.2 | 313 | 10\% |  |  |  |

Table 13. Bss.27.4.bc, 7.a, d-h: Estimates of recreational catches of seabass in different countries and years in numbers and weight of fish for retained and released components of the catch, and release rates. The relative standard error (RSE) is provided where available and expressed as a percentage.

| Country | Year | Area | Numbers (thousands) <br> Retained | Weight (tonnes) RSE | Released | RSE | Total | RSE | \% released | Retained | RSE | Released | RSE | Total | RSE | \% released | Source |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Belgium | 2012 | $\begin{aligned} & \text { BSS- } \\ & 47 \end{aligned}$ |  |  |  |  |  |  |  | 60 |  |  |  |  |  |  | Unknown |
| France | 2009-2011 | $\begin{aligned} & \text { BSS- } \\ & 47 \end{aligned}$ | 781 |  | 796 |  | 1578 | >26 | 50 | 940 |  | 332 |  | 1272 | >26 | 26 | ICES (2014b) |
|  | 2009-2011 | $\begin{aligned} & \text { BSS- } \\ & \text { 8AB } \end{aligned}$ | 1168 |  | 1190 |  | 2357 | >26 | 50 | 1405 |  | 496 |  | 1901 | >26 | 26 | Calculated |
|  | 2009-2011 | Both | 1949 |  | 1986 |  | 3935 | 26 | 50 | 2345 |  | 828 |  | 3173 | 26 | 26 | Rocklin et al. (2014) |
|  | 2011-2012 | $\begin{aligned} & \text { BSS- } \\ & 47 \end{aligned}$ | 2043 |  | 1581 |  | 3624 |  | 44 | 2458 |  | 659 |  | 3117 |  | 21 | Ifremer |
|  | 2011-2012 | $\begin{aligned} & \text { BSS- } \\ & \text { 8AB } \end{aligned}$ | 572 |  | 281 |  | 852 |  | 33 | 688 |  | 117 |  | 805 |  | 15 | Ifremer |
|  | 2011-2012 | All | 2615 |  | 1861 |  | 3935 |  | 47 | 3146 |  | 776 |  | 3922 |  | 20 | Ifremer |
| Netherlands | 2010-2011 | $\begin{aligned} & \text { BSS- } \\ & 47 \end{aligned}$ | 234 | 38 | 131 | 27 | 366 | 30 | 36 | 138 | 37 |  |  |  |  |  | van der Hammen and de Graaf (2013) |
|  | 2012-2013 | $\begin{aligned} & \text { BSS- } \\ & 47 \end{aligned}$ | 335 | 26 | 332 | 21 | 667 |  | 50 | 229 | 26 |  |  |  |  |  | van der Hammen and de Graaf (2015) |
|  | 2014-2015 | $\begin{aligned} & \text { BSS- } \\ & 47 \end{aligned}$ | 176 | 19 | 499 | 20 | 675 |  | 74 | 138 | 20 |  |  |  |  |  | van der Hammen and de Graaf (2017) |
| UK | 2012-2013 | $\begin{aligned} & \text { BSS- } \\ & 47 \end{aligned}$ | 367 |  | 576 |  | 943 |  | 61 | $\begin{aligned} & 230- \\ & 440 \end{aligned}$ |  | $\begin{aligned} & 150- \\ & 250 \end{aligned}$ |  | $\begin{aligned} & 380- \\ & 690 \end{aligned}$ | $\begin{aligned} & 26- \\ & 38 \end{aligned}$ | 36-39 | Armstrong et al. (2013) |

Table 14. Bss.27.4.bc, 7.a, d-h: Recreational removals (tonnes) by country for 2012. PRM indicates fish that die after release, applying post release mortality of 5\% as used in the WKBASS assessment WK 2018.

| Country | Year | Retained | PRM | Removals |
| :--- | :---: | :---: | :---: | :---: | :---: |
| France | $2009-2011$ | 940 | 17 | 957 |
| Netherlands | $2010-2011$ | 138 | 3 | 141 |
| England | 2012 | 332 | 10 | 343 |
| Total | 2012 | 1410 | 29 | 1440 |

Table 15. Bss.27.4.bc, 7.a, $d-h$ : Values of expected recreational F reductions associated with management measures applied to Bss.27.4bc7ad-h since 2015. Frec multiplier represents the recreational F relative to 2012.

| Mear | MANAGEMENT SCE- <br> NARIO | Frec Multiplier |  |  |
| :--- | :--- | :--- | :--- | :--- |
|  | MCRS | Bag limit | Open season |  |
| Pre-2015 | 36 cm | none | none | 1.000 |
| 2015 | 42 cm for 6 months | three-fish for 6 <br> months | none | 0.832 |
| $2016 \& 2017$ | 42 cm | one fish | 6 months | 0.282 |
| 2018 | 42 cm | no fish | 3 months | 0.191 |

Table 16. Bss.27.4.bc, 7.a, d-h: Country specific proportion reduction in retained catch numbers obtained by applying bag limits and increased MRS from 36 to $\mathbf{4 2} \mathbf{~ c m}$ to catch numbers in fishing trips observed in national recreational fishing surveys taking place before the new management measures were introduced (Armstrong et al., 2014). The mean weights in kg of retained and released fish from surveys are shown. BL represents bag limit and MCRS is the increase from 36 cm to 42 cm .

|  | Manage- <br> ment <br> measure | Weights <br> (kg) | BL2 |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Country | BL1 | BL2 | BL3 |  | MCRS only | Retained | Released |  |  |
| France (all) | 0.61 | 0.46 | 0.39 | 0.36 | 0.35 | 0.35 | 1.20 | 0.42 |  |
| Nether- <br> lands | 0.64 | 0.64 | 0.64 | 0.64 | 0.64 | 0.64 | 0.59 | $0.40^{*}$ |  |
| UK | 0.52 | 0.32 | 0.23 | 0.23 | 0.23 | 0.23 | 1.09 | 0.39 |  |

[^18]Table 17. Bss.27.4.bc, 7.a, d-h: Values of expected recreational F reductions associated with management measures applied to Bss.27.4bc7ad-h reflecting any combination of bag limit (BL) and open season length (months).

| Open Season | BL 1 fish | BL 2 fish | BL 3 fish | BL 4 fish | BL 5 fish | No BL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.099 | 0.099 | 0.099 | 0.099 | 0.099 | 0.099 |
| 3 | 0.191 | 0.221 | 0.235 | 0.239 | 0.240 | 0.240 |
| 6 | 0.282 | 0.343 | 0.371 | 0.379 | 0.381 | 0.381 |
| 7 | 0.312 | 0.383 | 0.416 | 0.426 | 0.428 | 0.428 |
| 10 | 0.404 | 0.464 | 0.506 | 0.519 | 0.522 | 0.522 |
| 12 | 0.465 | 0.586 | 0.552 | 0.566 | 0.569 | 0.569 |

Table 18. Estimated length-based maturity ogive parameters.

|  | Females | Males |
| :--- | :---: | :---: |
| Intercept (a) | -13.556 | -16.851 |
| Slope (b) | 0.3335 | 0.4861 |
| c =a/b | -40.6488 | -34.6652 |
| L25\% | 37.35 | 32.41 |
| L50\% | 40.65 | 34.67 |
| L75\% | 43.95 | 36.93 |

Table 19. Proportion mature-at-age (females) derived by Stock Synthesis model.

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Pmat | 0.000 | 0.000 | 0.000 | 0.000 | 0.186 | 0.419 | 0.638 | 0.792 | 0.885 | 0.937 |
| Age | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | $19+$ |
| Pmat | 0.965 | 0.980 | 0.989 | 0.993 | 0.996 | 0.998 | 0.998 | 0.999 | 0.999 | 1.000 |

Table 20. Inferences on natural mortality rate from a range of life-history based methods (WKBASS Data WK update of table provided by ICES, IBPNEW 2012 seabass benchmark; see data WK report for full list of references).

| Source <br> Hoenig 1983 | Formulation <br> variety of taxa $\ln (M)=1.44-0.982 * \ln (\operatorname{tmax})$; <br> teleosts $\quad \ln (M)=1.46-1.01 * \ln (\operatorname{tmax})$ | Combined sex M |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | tmax 28 | tmax 27 | tmax 26 |
|  |  | 0.160 | 0.166 | 0.160 |
|  |  | 0.149 | 0.154 | 0.160 |
| Then et al 2015 | $\mathrm{M}=4.899 * \operatorname{tmax}^{\wedge}-916$ (from 226 species) | 0.231 | 0.239 | 0.248 |
|  | $\mathrm{M}=4.118{ }^{*} \mathrm{~K}^{\wedge} 0.73$. $\mathrm{Linf}^{\wedge}-0.33$ | 0.173 |  |  |
| Alverson and Carney 1975 | $\mathrm{M}=3 \mathrm{k} /\left(\exp \left(0.38^{*} \mathrm{tmax}^{*} \mathrm{k}\right)-1\right)$ | 0.161 | 0.171 | 0.181 |
| Pauly 1980 | $M=\exp \left(-0.0152+0.6543^{*} \ln (\mathrm{k})-0.279 * \ln (\mathrm{Linf}, \mathrm{cm})+0.4634 * \operatorname{lnT}(\mathrm{OC})\right)$ | 0.196 | TdegC= | 12 |
|  |  | 0.211 | TdegC= | 14 |
|  |  | 0.224 | TdegC= | 16 |
| Ralston 1987 | $\mathrm{M}=0.0189+2.06 * \mathrm{k}$ | 0.219 |  |  |
| Beverton 1992 | $\mathrm{M}=3 \mathrm{k} /\left(\exp \left(a m^{*} \mathrm{k}\right)-1\right) \quad \mathrm{am}=$ age at $50 \%$ maturity | $\begin{aligned} & 0.369 \text { female } a m \text {; comb sex k } \\ & 0.614 \text { male } a m, \text { comb sex k } \\ & \hline \end{aligned}$ |  |  |
|  |  |  |  |  |
| Jensen (1997) | $\mathrm{M}=1.5 \mathrm{~K}$ | 0.146 |  |  |
| Gislason 2010 <br> Lorenzen | $\begin{aligned} & M=\exp \left(0.55-1.61^{*} \operatorname{Ln}(L)+1.44^{*} \operatorname{Ln}(\operatorname{Linf})+\operatorname{Ln}(K)\right) \\ & M=3^{*} W^{\wedge}-0.288 \end{aligned}$ | age 1age 3age 5 | Gislason | Lorenzen |
|  |  |  | 1.599 | 1.210 |
|  |  |  | 0.539 | 0.644 |
|  |  |  | 0.312 | 0.482 |
|  | Gislason: L = length at age from VBGF | age 7 | 0.221 | 0.402 |
|  | Lorenzen: W = mean wt at age from 2016 WGCSE SS3 run | age 9 | 0.175 | 0.355 |
|  |  | age 15 | 0.117 | 0.287 |
|  |  | age 20 | 0.100 | 0.262 |


| Life history parameters |  |
| :--- | :--- |
| VBGF K (combined sex) | 0.097 |
| VBGF Linf (combined sex) | 84.55 |
| VBGF to (combined sex | -0.73 |
| Age at $50 \%$ maturity females (L50\% converted to age) | 6 |
| Age at $50 \%$ maturity males (L50\% converted to age) | 4 |
|  |  |
| Max age (combined sex) | 28 |
| Length at $50 \%$ mat females | 40.65 |
| Length at $50 \%$ mat males | 34.67 |

Table 21. Bss-47: Inferences on natural mortality rate by age class using the Gislason et al. (2010) and Lorenzen (2006) methods. Values are given unscaled, and scaled to a mean M of 0.24 at ages 10-20 (based on Then et al., 2015 for maximum age of 27 years) and mean $M$ of 0.15 at ages 10-20 (from Hoenig, 1983 using maximum age of 27-28 years).

|  | Gislason method M | Lorenzen method M |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age class | L | Not scaled | Scaled to <br> 0.24 at ages 10-20 | Scaled to <br> 0.15 at age $5-20$ | W <br> (kg) | Not scaled | Scaled to <br> 0.24 at ages 10-20 | Scaled to <br> 0.15 at age 5-20 |
| 1 | 13.1 | 1.599 | 3.145 | 1.966 | 0.023 | 1.210 | 0.995 | 0.622 |
| 2 | 19.7 | 0.827 | 1.627 | 1.017 | 0.096 | 0.807 | 0.663 | 0.415 |
| 3 | 25.7 | 0.539 | 1.060 | 0.662 | 0.209 | 0.644 | 0.530 | 0.331 |
| 4 | 31.1 | 0.395 | 0.778 | 0.486 | 0.369 | 0.547 | 0.450 | 0.281 |
| 5 | 36.1 | 0.312 | 0.613 | 0.383 | 0.570 | 0.482 | 0.397 | 0.248 |
| 6 | 40.5 | 0.258 | 0.508 | 0.317 | 0.807 | 0.436 | 0.359 | 0.224 |
| 7 | 44.6 | 0.221 | 0.435 | 0.272 | 1.073 | 0.402 | 0.331 | 0.207 |
| 8 | 48.3 | 0.195 | 0.383 | 0.239 | 1.359 | 0.376 | 0.309 | 0.193 |
| 9 | 51.6 | 0.175 | 0.344 | 0.215 | 1.659 | 0.355 | 0.292 | 0.182 |
| 10 | 54.7 | 0.159 | 0.314 | 0.196 | 1.968 | 0.338 | 0.278 | 0.174 |
| 11 | 57.5 | 0.147 | 0.290 | 0.181 | 2.279 | 0.324 | 0.266 | 0.166 |
| 12 | 60.0 | 0.138 | 0.270 | 0.169 | 2.588 | 0.312 | 0.257 | 0.160 |
| 13 | 62.2 | 0.130 | 0.255 | 0.159 | 2.893 | 0.302 | 0.249 | 0.155 |
| 14 | 64.3 | 0.123 | 0.242 | 0.151 | 3.190 | 0.294 | 0.242 | 0.151 |
| 15 | 66.2 | 0.117 | 0.231 | 0.144 | 3.476 | 0.287 | 0.236 | 0.147 |
| 16 | 67.9 | 0.113 | 0.222 | 0.138 | 3.751 | 0.280 | 0.231 | 0.144 |
| 17 | 69.4 | 0.109 | 0.214 | 0.134 | 4.013 | 0.275 | 0.226 | 0.141 |
| 18 | 70.8 | 0.105 | 0.207 | 0.129 | 4.262 | 0.270 | 0.222 | 0.139 |
| 19 | 72.1 | 0.102 | 0.201 | 0.126 | 4.498 | 0.266 | 0.219 | 0.137 |
| 20 | 73.2 | 0.100 | 0.196 | 0.122 | 4.719 | 0.262 | 0.216 | 0.135 |
| 21 | 74.3 | 0.097 | 0.192 | 0.120 | 4.926 | 0.259 | 0.213 | 0.133 |
| 22 | 75.2 | 0.095 | 0.188 | 0.117 | 5.119 | 0.256 | 0.211 | 0.132 |
| 23 | 76.1 | 0.094 | 0.184 | 0.115 | 5.299 | 0.254 | 0.209 | 0.130 |
| 24 | 76.9 | 0.092 | 0.181 | 0.113 | 5.464 | 0.252 | 0.207 | 0.129 |
| 25 | 77.6 | 0.091 | 0.179 | 0.112 | 5.616 | 0.250 | 0.205 | 0.128 |
| 26 | 78.2 | 0.090 | 0.176 | 0.110 | 5.755 | 0.248 | 0.204 | 0.127 |
| 27 | 78.8 | 0.089 | 0.174 | 0.109 | 5.882 | 0.246 | 0.203 | 0.127 |
| 28 | 79.3 | 0.088 | 0.172 | 0.108 | 5.996 | 0.245 | 0.201 | 0.126 |
| mean over $\begin{gathered} \text { ages } 10- \\ 20 \end{gathered}$ |  | 0.122 | 0.240 | 0.150 | 3.422 | 0.292 | 0.240 | 0.150 |

Table 22. Updated time-series of Cefas Solent autumn survey of juvenile seabass, including 2013 survey results. Indices for 2000 are revised. A change in trawl design took place in 1993, and calibration factors are applied.


Table 23. Seabass indices of abundance 2000-2014 (swept area) from the Channel Groundfish Survey. The relative standard error CV is the log-transformed value used in SS3 (sqrt(loge (1+CV^2)).

| year | Total <br> hauls | No. hauls with seabass | Percentage of hauls with seabass | Mean no. seabass per positive haul | Swept-area abundance index | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 68 | 6 | 9 | 2 | 245776 | 0.15 |
| 1989 | 61 | 3 | 5 | 1 | 77716 | 0.58 |
| 1990 | 75 | 8 | 11 | 8 | 1129914 | 0.12 |
| 1991 | 79 | 19 | 24 | 9 | 4250636 | 0.03 |
| 1992 | 60 | 23 | 38 | 13 | 2617986 | 0.11 |
| 1993 | 65 | 21 | 32 | 8 | 2299919 | 0.10 |
| 1994 | 86 | 19 | 22 | 5 | 1097828 | 0.11 |
| 1995 | 166 | 17 | 10 | 5 | 1021741 | 0.09 |
| 1996 | 134 | 26 | 19 | 3 | 1224238 | 0.13 |
| 1997 | 169 | 31 | 18 | 6 | 1817599 | 0.12 |
| 1998 | 82 | 38 | 46 | 8 | 2531043 | 0.08 |
| 1999 | 102 | 37 | 36 | 8 | 1642271 | 0.12 |
| 2000 | 100 | 36 | 36 | 9 | 2570994 | 0.08 |
| 2001 | 109 | 39 | 36 | 9 | 3150674 | 0.14 |
| 2002 | 100 | 44 | 44 | 12 | 3872427 | 0.11 |
| 2003 | 94 | 41 | 44 | 20 | 8739056 | 0.11 |
| 2004 | 94 | 44 | 47 | 8 | 3598436 | 0.10 |
| 2005 | 105 | 40 | 38 | 7 | 3005315 | 0.08 |
| 2006 | 110 | 36 | 33 | 14 | 5518000 | 0.12 |
| 2007 | 103 | 33 | 32 | 8 | 3661314 | 0.14 |
| 2008 | 105 | 40 | 38 | 10 | 6468839 | 0.15 |
| 2009 | 102 | 26 | 26 | 7 | 2564694 | 0.09 |
| 2010 | 101 | 30 | 30 | 4 | 1804538 | 0.10 |
| 2011 | 108 | 27 | 25 | 4 | 1513742 | 0.12 |
| 2012 | 96 | 25 | 26 | 5 | 2034552 | 0.11 |
| 2013 | 96 | 19 | 20 | 4 | 995987 | 0.13 |
| 2014 | 98 | 20 | 20 | 3 | 669931 | 0.13 |

Table 24. Numbers-at-age in Solent Survey1986-2015: updated time-series of Cefas Solent autumn survey of juvenile seabass (2015 revised).

| AGE CLASS | 2 | 3 | 4 |
| :---: | :---: | :---: | :---: |
| 1986 | 0.27 | 4.26 | 1.31 |
| 1987 | 0.05 | 0.28 | 2.27 |
| 1989 | 6.68 | 0.37 | 0 |
| 1990 | 2.81 | 1.15 | 0.02 |
| 1991 | 3.1 | 0.2 | 0 |
| 1992 | 0.95 | 18.6 | 0.16 |
| 1993 | 6.65 | 3.59 | 4.39 |
| 1994 | 3.33 | 1.84 | 0.29 |
| 1995 | 4.83 | 4.69 | 0.72 |
| 1996 | 5.52 | 0.43 | 0.11 |
| 1997 | 33.6 | 4.52 | 0.06 |
| 1998 | 1.22 | 5.5 | 0.61 |
| 1999 | 19.4 | 0.67 | 0.87 |
| 2000 | 6.07 | 11.4 | 0.03 |
| 2001 | 34.4 | 3.92 | 1.57 |
| 2002 | 7.42 | 3.87 | 0.4 |
| 2003 | 8.37 | 4.6 | 0.59 |
| 2005 | 13.1 | 7.98 | 0.84 |
| 2006 | 9.51 | 9.21 | 1.02 |
| 2007 | 3.42 | 1.78 | 0.3 |
| 2008 | 18.5 | 6.66 | 0.34 |
| 2009 | 13.3 | 6.25 | 0.33 |
| 2011 | 2.3 | 1.4 | 0.4 |
| 2013 | 1.34 | 0.08 | 0.1 |
| 2014 | 1.17 | 1.02 | 0.11 |
| 2015 | 6.95 | 0.44 | 0.05 |
| 2016 | 3.75 | 2.17 | 0.11 |
| 2017 | 0.858 | 2.562 | 0.118 |
| 2018 | 2.168 | 0.316 | 0.18 |

Table 25. Key model assumptions and parameters from the WGCSE 2019 update assessment.

| Characteristic | Settings |
| :---: | :---: |
| Starting year | 1985 |
| Ending year | Assessment year-1 (2018) |
| Equilibrium commercial catch for starting year | 0.82* landings in 1985 by fleet. |
| Equilibrium recreational catch for starting year | Constant $F$ estimated using 2012 survey results 19852014; 2015-present Frec multiplier on F 2012 survey results |
| Number of areas | 1 |
| Number of seasons | 1 |
| Number of fishing fleets | 6 |
| Number of surveys | 2: CGFS; Solent autumn survey. |
| Number of commercial tunning fleets | 1 |
| Individual growth | von Bertalanffy, parameters fixed, combined sex |
| Number of active parameters | 111 |
| Population characteristics |  |
| Maximum age | 30 |
| Genders | 1 |
| Population length bins | 4-100, 2 cm bins |
| Ages for summary total biomass | 0-30 |
| Data characteristics |  |
| Data length bins (for length structured fleets) | 6-94, 2 cm bins |
| Data age bins (for age structured fleets) | 0-16+ |
| Minimum age for growth model | 2 |
| Maximum age for growth model | 30 |
| Maturity | Logistic 2-parameter - females; L50 = 40.65 cm |
| Fishery characteristics |  |
| Fishery timing | -1 (whole year) |
| Fishing mortality method | Hybrid |
| Maximum F | 2.9 |
| Fleet 1: UK Trawl/nets selectivity | Double normal, length-based |
| Fleet 2: UK Line selectivity | Asymptotic, length-based |
| Fleet 3: UK Midwater trawl selectivity | Asymptotic, length-based |
| Fleet 4: Combined French fleet selectivity | Asymptotic 1985-2014, Double normal 2015-present, length-based |
| Fleet 5: Other fleets/gears selectivity | Mirrors French fleet |
| Fleet 6: Rrecreational fishery | Double normal, length-based |
| Blocks: Selectivity and Retention | Fleets 1, 2, 4, 5 and 62015 to present |
| Survey characteristics |  |
| Solent autumn survey timing (yr) | 0.83 |
| CGFS survey timing (yr) | 0.75 |


| Characteristic | Settings |
| :---: | :---: |
| French LPUE timing (yr) | -1 |
| Catchabilities (all surveys) | Analytical solution |
| Survey selectivities: Solent autumn: | Double normal, length-based constrained by Min-Max age selectivity, age-based |
| Survey selectivities: CGFS | Double normal, length-based |
| Tunning fleet: French LPUE | Mirrors French fleet |
| Fixed biological characteristics |  |
| Natural mortality | 0.24 |
| Beverton-Holt steepness | 0.999 |
| Recruitment variability ( $\sigma \mathrm{R}$ ) | 0.9 |
| Weight-length coefficient | 0.00001296 |
| Weight-length exponent | 2.969 |
| Maturity inflection (L50\%) | 40.649 cm |
| Maturity slope | -0.33349 |
| Length-at-age Amin | 19.6 cm at $\mathrm{Amin}=2$ |
| Length-at-Amax | 80.26 cm |
| von Bertalanffy k | 0.09699 |
| von Bertalanffy Linf | 84.55 cm |
| von Bertalanffy t0 | -0.730 yr |
| Std. Deviation length-at-age (cm) | SD $=0.1166$ * age + 3.5609 |
| Age error matrix | CV 12\% at-age |
| Other model settings |  |
| First year for main recruitment deviations for burn-in period | 1955 |
| Last year for recruit deviations | 2016 (last year class with survey indices-2) |
| Last year no bias adjustment | 1974.5 |
| First year full bias adjustment | 1881.7 |
| Last year full bias adjustment | 2014.9 |
| First year recent year no bias adjustment | 2016.8 |
| Maximum bias adjustment | 0.907 |

Table 26. Final seabass update assessment: model estimated stock numbers-at-age (thousands of fish).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 862 | 1258 | 20430 | 8597 | 4948 | 1675 | 1524 | 1302 | 1794 | 4213 | 1581 | 984 | 722 | 549 | 401 | 278 | 619 |
| 1986 | 2469 | 678 | 988 | 15984 | 6658 | 3764 | 1245 | 1108 | 932 | 1271 | 2971 | 1112 | 692 | 507 | 386 | 282 | 631 |
| 1987 | 20458 | 1943 | 533 | 772 | 12343 | 5035 | 2772 | 894 | 780 | 648 | 878 | 2047 | 765 | 476 | 349 | 266 | 629 |
| 1988 | 16119 | 16093 | 1525 | 415 | 593 | 9230 | 3642 | 1941 | 610 | 523 | 430 | 580 | 1349 | 504 | 314 | 230 | 590 |
| 1989 | 93569 | 12680 | 12640 | 1191 | 320 | 447 | 6767 | 2600 | 1357 | 421 | 359 | 294 | 396 | 922 | 345 | 215 | 561 |
| 1990 | 7374 | 73604 | 9959 | 9874 | 918 | 241 | 327 | 4823 | 1816 | 936 | 289 | 245 | 201 | 271 | 630 | 236 | 531 |
| 1991 | 14562 | 5800 | 57809 | 7779 | 7609 | 691 | 176 | 233 | 3361 | 1249 | 640 | 197 | 167 | 137 | 185 | 430 | 524 |
| 1992 | 22786 | 11455 | 4554 | 45103 | 5976 | 5689 | 500 | 124 | 159 | 2264 | 835 | 427 | 131 | 111 | 91 | 123 | 637 |
| 1993 | 10557 | 17924 | 8993 | 3552 | 34645 | 4471 | 4117 | 350 | 84 | 107 | 1509 | 554 | 283 | 87 | 74 | 61 | 505 |
| 1994 | 32088 | 8305 | 14073 | 7017 | 2731 | 25966 | 3244 | 2896 | 241 | 57 | 72 | 1009 | 370 | 189 | 58 | 49 | 378 |
| 1995 | 51030 | 25241 | 6523 | 10995 | 5407 | 2052 | 18906 | 2297 | 2011 | 165 | 39 | 49 | 688 | 252 | 129 | 40 | 293 |
| 1996 | 2914 | 40142 | 19823 | 5092 | 8453 | 4045 | 1484 | 13263 | 1577 | 1363 | 111 | 26 | 33 | 462 | 170 | 87 | 225 |
| 1997 | 56168 | 2292 | 31495 | 15424 | 3886 | 6238 | 2859 | 1006 | 8702 | 1012 | 865 | 70 | 16 | 21 | 291 | 107 | 197 |
| 1998 | 18334 | 44183 | 1799 | 24526 | 11791 | 2876 | 4432 | 1954 | 667 | 5663 | 652 | 554 | 45 | 11 | 13 | 187 | 195 |
| 1999 | 54392 | 14422 | 34677 | 1402 | 18775 | 8754 | 2053 | 3044 | 1303 | 437 | 3666 | 420 | 357 | 29 | 7 | 9 | 246 |
| 2000 | 27423 | 42786 | 11318 | 27004 | 1072 | 13905 | 6213 | 1396 | 2002 | 839 | 278 | 2323 | 266 | 226 | 18 | 4 | 162 |
| 2001 | 29225 | 21572 | 33579 | 8815 | 20668 | 796 | 9939 | 4278 | 933 | 1313 | 544 | 179 | 1497 | 171 | 145 | 12 | 107 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 47319 | 22990 | 16930 | 26156 | 6748 | 15351 | 569 | 6840 | 2858 | 611 | 851 | 351 | 116 | 963 | 110 | 94 | 77 |
| 2003 | 45959 | 37223 | 18044 | 13190 | 20022 | 5010 | 10964 | 391 | 4571 | 1875 | 397 | 550 | 227 | 75 | 623 | 71 | 110 |
| 2004 | 34346 | 36153 | 29200 | 14031 | 10051 | 14738 | 3529 | 7392 | 255 | 2908 | 1177 | 248 | 343 | 141 | 46 | 388 | 113 |
| 2005 | 22710 | 27018 | 28358 | 22697 | 10682 | 7385 | 10349 | 2369 | 4788 | 161 | 1815 | 730 | 153 | 212 | 87 | 29 | 311 |
| 2006 | 25061 | 17864 | 21181 | 22000 | 17201 | 7781 | 5116 | 6821 | 1499 | 2947 | 98 | 1090 | 437 | 92 | 127 | 52 | 204 |
| 2007 | 27150 | 19714 | 14005 | 16432 | 16671 | 12522 | 5384 | 3369 | 4315 | 923 | 1784 | 59 | 653 | 262 | 55 | 76 | 153 |
| 2008 | 15775 | 21357 | 15460 | 10878 | 12486 | 12195 | 8729 | 3582 | 2160 | 2699 | 569 | 1093 | 36 | 399 | 160 | 34 | 141 |
| 2009 | 13076 | 12409 | 16750 | 12010 | 8265 | 9130 | 8500 | 5813 | 2302 | 1356 | 1671 | 350 | 671 | 22 | 245 | 98 | 107 |
| 2010 | 3423 | 10286 | 9734 | 13019 | 9137 | 6060 | 6392 | 5696 | 3765 | 1458 | 847 | 1038 | 217 | 416 | 14 | 152 | 128 |
| 2011 | 13192 | 2692 | 8064 | 7551 | 9858 | 6635 | 4174 | 4184 | 3581 | 2303 | 878 | 506 | 619 | 129 | 248 | 8 | 167 |
| 2012 | 5831 | 10377 | 2112 | 6263 | 5732 | 7194 | 4607 | 2763 | 2668 | 2228 | 1413 | 535 | 308 | 377 | 79 | 151 | 107 |
| 2013 | 16744 | 4587 | 8136 | 1638 | 4736 | 4146 | 4924 | 2995 | 1727 | 1624 | 1337 | 843 | 319 | 183 | 225 | 47 | 155 |
| 2014 | 18780 | 13171 | 3594 | 6297 | 1231 | 3385 | 2784 | 3118 | 1812 | 1013 | 937 | 765 | 481 | 182 | 105 | 129 | 116 |
| 2015 | 3661 | 14773 | 10334 | 2792 | 4763 | 887 | 2304 | 1808 | 1962 | 1120 | 621 | 573 | 468 | 295 | 112 | 65 | 152 |
| 2016 | 13452 | 2880 | 11614 | 8087 | 2145 | 3503 | 612 | 1500 | 1142 | 1228 | 700 | 389 | 361 | 296 | 187 | 71 | 139 |
| 2017 | 22031 | 10581 | 2265 | 9114 | 6284 | 1626 | 2552 | 429 | 1029 | 779 | 838 | 479 | 267 | 248 | 204 | 130 | 147 |
| 2018 | 22019 | 17330 | 8322 | 1778 | 7087 | 4779 | 1197 | 1825 | 302 | 722 | 546 | 588 | 337 | 188 | 175 | 144 | 196 |
| 2019 | 22020 | 17321 | 13631 | 6537 | 1389 | 5465 | 3596 | 876 | 1314 | 217 | 517 | 392 | 423 | 242 | 136 | 126 | 246 |

Table 27. Final seabass update assessment: model estimated fishing mortality-at-age.

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0.000 | 0.001 | 0.005 | 0.016 | 0.034 | 0.056 | 0.078 | 0.095 | 0.105 | 0.110 | 0.112 | 0.112 | 0.112 | 0.112 | 0.112 | 0.112 | 0.112 |
| 1986 | 0.000 | 0.001 | 0.007 | 0.019 | 0.039 | 0.066 | 0.091 | 0.111 | 0.123 | 0.130 | 0.133 | 0.134 | 0.134 | 0.134 | 0.134 | 0.133 | 0.133 |
| 1987 | 0.000 | 0.002 | 0.009 | 0.024 | 0.051 | 0.084 | 0.117 | 0.143 | 0.160 | 0.170 | 0.175 | 0.177 | 0.178 | 0.178 | 0.177 | 0.177 | 0.177 |
| 1988 | 0.000 | 0.002 | 0.007 | 0.020 | 0.042 | 0.070 | 0.097 | 0.117 | 0.130 | 0.137 | 0.140 | 0.140 | 0.141 | 0.140 | 0.140 | 0.139 | 0.138 |
| 1989 | 0.000 | 0.002 | 0.007 | 0.020 | 0.043 | 0.072 | 0.099 | 0.119 | 0.131 | 0.138 | 0.140 | 0.141 | 0.141 | 0.141 | 0.140 | 0.140 | 0.138 |
| 1990 | 0.000 | 0.002 | 0.007 | 0.021 | 0.044 | 0.073 | 0.100 | 0.121 | 0.134 | 0.140 | 0.143 | 0.144 | 0.144 | 0.143 | 0.142 | 0.142 | 0.141 |
| 1991 | 0.000 | 0.002 | 0.008 | 0.024 | 0.051 | 0.085 | 0.117 | 0.140 | 0.155 | 0.163 | 0.166 | 0.167 | 0.166 | 0.166 | 0.165 | 0.164 | 0.163 |
| 1992 | 0.000 | 0.002 | 0.009 | 0.024 | 0.050 | 0.083 | 0.115 | 0.140 | 0.157 | 0.166 | 0.170 | 0.171 | 0.171 | 0.171 | 0.171 | 0.170 | 0.169 |
| 1993 | 0.000 | 0.002 | 0.008 | 0.023 | 0.048 | 0.081 | 0.112 | 0.136 | 0.151 | 0.159 | 0.163 | 0.164 | 0.164 | 0.164 | 0.164 | 0.163 | 0.162 |
| 1994 | 0.000 | 0.001 | 0.007 | 0.021 | 0.046 | 0.077 | 0.105 | 0.125 | 0.136 | 0.142 | 0.143 | 0.143 | 0.143 | 0.142 | 0.141 | 0.140 | 0.138 |
| 1995 | 0.000 | 0.002 | 0.008 | 0.023 | 0.050 | 0.084 | 0.115 | 0.136 | 0.149 | 0.155 | 0.158 | 0.158 | 0.157 | 0.156 | 0.155 | 0.154 | 0.152 |
| 1996 | 0.000 | 0.003 | 0.011 | 0.030 | 0.064 | 0.107 | 0.149 | 0.181 | 0.203 | 0.215 | 0.220 | 0.222 | 0.223 | 0.222 | 0.222 | 0.221 | 0.219 |
| 1997 | 0.000 | 0.002 | 0.010 | 0.029 | 0.061 | 0.102 | 0.141 | 0.171 | 0.190 | 0.200 | 0.205 | 0.206 | 0.206 | 0.205 | 0.204 | 0.203 | 0.201 |
| 1998 | 0.000 | 0.002 | 0.010 | 0.027 | 0.058 | 0.097 | 0.136 | 0.165 | 0.185 | 0.195 | 0.200 | 0.201 | 0.201 | 0.201 | 0.200 | 0.200 | 0.198 |
| 1999 | 0.000 | 0.002 | 0.010 | 0.028 | 0.060 | 0.103 | 0.146 | 0.179 | 0.200 | 0.211 | 0.216 | 0.218 | 0.218 | 0.218 | 0.217 | 0.216 | 0.215 |
| 2000 | 0.000 | 0.002 | 0.010 | 0.027 | 0.058 | 0.096 | 0.133 | 0.163 | 0.182 | 0.193 | 0.198 | 0.200 | 0.200 | 0.200 | 0.199 | 0.199 | 0.197 |
| 2001 | 0.000 | 0.002 | 0.010 | 0.027 | 0.057 | 0.096 | 0.134 | 0.164 | 0.183 | 0.194 | 0.198 | 0.200 | 0.200 | 0.200 | 0.199 | 0.199 | 0.197 |
| 2002 | 0.000 | 0.002 | 0.010 | 0.027 | 0.058 | 0.097 | 0.134 | 0.163 | 0.182 | 0.191 | 0.196 | 0.197 | 0.197 | 0.196 | 0.196 | 0.195 | 0.192 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0.000 | 0.003 | 0.012 | 0.032 | 0.066 | 0.110 | 0.154 | 0.189 | 0.212 | 0.225 | 0.231 | 0.233 | 0.234 | 0.233 | 0.233 | 0.232 | 0.231 |
| 2004 | 0.000 | 0.003 | 0.012 | 0.033 | 0.068 | 0.114 | 0.158 | 0.194 | 0.218 | 0.231 | 0.238 | 0.240 | 0.240 | 0.240 | 0.239 | 0.239 | 0.237 |
| 2005 | 0.000 | 0.003 | 0.014 | 0.037 | 0.077 | 0.127 | 0.177 | 0.218 | 0.245 | 0.262 | 0.270 | 0.273 | 0.274 | 0.274 | 0.273 | 0.273 | 0.272 |
| 2006 | 0.000 | 0.003 | 0.014 | 0.037 | 0.077 | 0.128 | 0.178 | 0.218 | 0.245 | 0.262 | 0.270 | 0.273 | 0.274 | 0.273 | 0.273 | 0.272 | 0.271 |
| 2007 | 0.000 | 0.003 | 0.013 | 0.035 | 0.073 | 0.121 | 0.167 | 0.204 | 0.229 | 0.243 | 0.250 | 0.253 | 0.253 | 0.253 | 0.252 | 0.251 | 0.250 |
| 2008 | 0.000 | 0.003 | 0.013 | 0.035 | 0.073 | 0.121 | 0.167 | 0.202 | 0.226 | 0.239 | 0.246 | 0.248 | 0.248 | 0.247 | 0.246 | 0.245 | 0.243 |
| 2009 | 0.000 | 0.003 | 0.012 | 0.033 | 0.070 | 0.117 | 0.160 | 0.194 | 0.217 | 0.230 | 0.236 | 0.238 | 0.238 | 0.237 | 0.236 | 0.235 | 0.233 |
| 2010 | 0.000 | 0.003 | 0.014 | 0.038 | 0.080 | 0.133 | 0.184 | 0.224 | 0.252 | 0.267 | 0.275 | 0.278 | 0.278 | 0.277 | 0.277 | 0.276 | 0.274 |
| 2011 | 0.000 | 0.003 | 0.013 | 0.036 | 0.075 | 0.125 | 0.173 | 0.210 | 0.235 | 0.248 | 0.255 | 0.257 | 0.257 | 0.256 | 0.255 | 0.254 | 0.252 |
| 2012 | 0.000 | 0.003 | 0.014 | 0.039 | 0.084 | 0.139 | 0.191 | 0.230 | 0.256 | 0.271 | 0.277 | 0.279 | 0.278 | 0.277 | 0.276 | 0.274 | 0.271 |
| 2013 | 0.000 | 0.004 | 0.016 | 0.045 | 0.096 | 0.158 | 0.217 | 0.263 | 0.293 | 0.310 | 0.318 | 0.321 | 0.320 | 0.319 | 0.317 | 0.316 | 0.313 |
| 2014 | 0.000 | 0.003 | 0.013 | 0.039 | 0.088 | 0.145 | 0.192 | 0.223 | 0.242 | 0.250 | 0.252 | 0.250 | 0.247 | 0.244 | 0.241 | 0.238 | 0.232 |
| 2015 | 0.000 | 0.001 | 0.005 | 0.024 | 0.067 | 0.131 | 0.189 | 0.220 | 0.229 | 0.229 | 0.226 | 0.223 | 0.219 | 0.215 | 0.211 | 0.208 | 0.202 |
| 2016 | 0.000 | 0.000 | 0.002 | 0.012 | 0.037 | 0.077 | 0.116 | 0.137 | 0.143 | 0.142 | 0.140 | 0.137 | 0.135 | 0.132 | 0.129 | 0.126 | 0.121 |
| 2017 | 0.000 | 0.000 | 0.002 | 0.012 | 0.034 | 0.066 | 0.095 | 0.110 | 0.114 | 0.114 | 0.113 | 0.112 | 0.110 | 0.109 | 0.107 | 0.106 | 0.102 |
| 2018 | 0.000 | 0.000 | 0.001 | 0.007 | 0.020 | 0.044 | 0.073 | 0.088 | 0.093 | 0.093 | 0.092 | 0.090 | 0.089 | 0.087 | 0.086 | 0.084 | 0.082 |

Table 28. Final seabass update assessment: stock summary table.

| Year | Recruitment (Age 0, thousands) | High | Low | SSB (Tonnes) | High | Low | F(4-15) | High | Low | Commercial landings | Commercial discards* | Recreational removals |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 862 | 1649 | 74 | 24810 | 31019 | 18600 | 0.096 | 0.122 | 0.07 | 994 |  | 1713 |
| 1986 | 2469 | 4506 | 433 | 21938 | 27485 | 16390 | 0.113 | 0.143 | 0.083 | 1318 |  | 1550 |
| 1987 | 20458 | 26382 | 14534 | 19651 | 24584 | 14719 | 0.149 | 0.188 | 0.11 | 1979 |  | 1412 |
| 1988 | 16119 | 23124 | 9114 | 17849 | 22272 | 13426 | 0.119 | 0.15 | 0.089 | 1239 |  | 1305 |
| 1989 | 93569 | 107559 | 79578 | 17065 | 21150 | 12979 | 0.12 | 0.151 | 0.09 | 1161 |  | 1204 |
| 1990 | 7374 | 12533 | 2215 | 15593 | 19422 | 11764 | 0.122 | 0.155 | 0.09 | 1064 |  | 1082 |
| 1991 | 14562 | 20522 | 8602 | 13634 | 17179 | 10088 | 0.142 | 0.181 | 0.103 | 1226 |  | 988 |
| 1992 | 22786 | 29774 | 15797 | 11894 | 15123 | 8665 | 0.145 | 0.183 | 0.106 | 1186 |  | 998 |
| 1993 | 10557 | 15661 | 5453 | 12117 | 15075 | 9160 | 0.139 | 0.171 | 0.107 | 1256 |  | 1150 |
| 1994 | 32088 | 41332 | 22843 | 14699 | 17500 | 11898 | 0.124 | 0.148 | 0.099 | 1370 |  | 1378 |
| 1995 | 51030 | 61438 | 40622 | 18556 | 21444 | 15668 | 0.136 | 0.161 | 0.111 | 1835 |  | 1551 |
| 1996 | 2914 | 5397 | 432 | 20666 | 23735 | 17597 | 0.188 | 0.22 | 0.153 | 3022 |  | 1570 |
| 1997 | 56168 | 67277 | 45059 | 19884 | 22991 | 16778 | 0.175 | 0.21 | 0.142 | 2620 |  | 1496 |
| 1998 | 18334 | 27674 | 8994 | 18670 | 21713 | 15627 | 0.17 | 0.2 | 0.138 | 2390 |  | 1448 |
| 1999 | 54392 | 66880 | 41905 | 18052 | 20997 | 15107 | 0.184 | 0.22 | 0.149 | 2670 |  | 1449 |
| 2000 | 27423 | 36891 | 17955 | 18133 | 21016 | 15250 | 0.169 | 0.2 | 0.136 | 2407 |  | 1496 |
| 2001 | 29225 | 41198 | 17253 | 19119 | 22062 | 16176 | 0.169 | 0.2 | 0.137 | 2500 |  | 1572 |
| 2002 | 47319 | 62144 | 32494 | 19937 | 22970 | 16904 | 0.167 | 0.198 | 0.135 | 2622 | 17 | 1655 |


| Year | Recruitment (Age 0, thousands) | High | Low | SSB (Tonnes) | High | Low | F(4-15) | High | Low | Commercial landings | Commercial discards* | Recreational removals |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 45959 | 58899 | 33020 | 21075 | 24227 | 17922 | 0.197 | 0.23 | 0.159 | 3459 | 16 | 1727 |
| 2004 | 34346 | 44995 | 23697 | 21776 | 25056 | 18495 | 0.2 | 0.24 | 0.163 | 3731 | 59 | 1771 |
| 2005 | 22710 | 30466 | 14954 | 22343 | 25749 | 18937 | 0.23 | 0.27 | 0.183 | 4430 | 96 | 1782 |
| 2006 | 25061 | 32603 | 17519 | 22078 | 25586 | 18571 | 0.23 | 0.27 | 0.182 | 4377 | 53 | 1777 |
| 2007 | 27150 | 35712 | 18588 | 21925 | 25455 | 18395 | 0.21 | 0.25 | 0.17 | 4064 | 50 | 1798 |
| 2008 | 15775 | 22582 | 8968 | 22552 | 26048 | 19057 | 0.21 | 0.25 | 0.168 | 4107 | 8 | 1824 |
| 2009 | 13076 | 18068 | 8083 | 23102 | 26555 | 19649 | 0.2 | 0.24 | 0.163 | 3889 | 151.2 | 1812 |
| 2010 | 3423 | 6296 | 549 | 22973 | 26363 | 19583 | 0.23 | 0.28 | 0.189 | 4562 | 147.9 | 1726 |
| 2011 | 13192 | 17646 | 8739 | 21351 | 24607 | 18095 | 0.22 | 0.26 | 0.175 | 3858 | 22 | 1592 |
| 2012 | 5831 | 8876 | 2787 | 19794 | 22871 | 16716 | 0.24 | 0.28 | 0.19 | 3987 | 156.6 | 1440 |
| 2013 | 16744 | 23902 | 9586 | 17758 | 20665 | 14851 | 0.27 | 0.33 | 0.21 | 4137 | 53.4 | 1241 |
| 2014 | 18780 | 28344 | 9215 | 14982 | 17763 | 12200 | 0.22 | 0.27 | 0.17 | 2682 | 24.7 | 1048 |
| 2015 | 3661 | 6468 | 854 | 12885 | 15595 | 10174 | 0.199 | 0.25 | 0.149 | 2066 | 39.5 | 737 |
| 2016 | 13452 | 24605 | 2298 | 11025 | 13689 | 8360 | 0.122 | 0.153 | 0.088 | 1295 | 198.6 | 228 |
| 2017 | 12383** |  |  | 10353 | 13020 | 7685 | 0.101 | 0.127 | 0.071 | 984 | 271.102 | 223 |
| 2018 | 12383** |  |  | 10313 | 13068 | 7559 | 0.079 | 0.101 | 0.055 | 801 | 482.4 | 156 |
| 2019 | 12383** |  |  |  |  |  |  |  |  |  |  |  |

*Partial discards, discard data are not available for all fleets in some years.
**Geometric mean recruitment 2005-2016.

Table 29. Inputs for short-term forecast. Fishing mortality is the estimates for 2015, which takes into account a change in overall selectivity due to the reduction in French landings. Numbers-at-ages 0-2 in 2015 are adjusted by replacing Stock Synthesis values for 0-group in 2014-2015 (years with no recruit deviations estimated) with the long-term GM, adjusted for natural mortality.

| age | 2018 | weight in stock | Pro- <br> portion <br> mature <br> (fe- <br> male) | H.Cons retained mean $F$ (2017) | H.Cons Discarded mean F (2017) | H.Cons retained mean weights | H.Cons discarded mean weights | H.Cons proportion retained | Recre-ational F | Recreational removals mean weight | M |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 12383 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.24 |
| 1 | 9741 | 0.024 | 0.000 | 0.000 | 0.000 | 0.103 | 0.103 | 0.304 | 0.000 | 0.077 | 0.24 |
| 2 | 7661 | 0.096 | 0.000 | 0.000 | 0.001 | 0.220 | 0.220 | 0.164 | 0.001 | 0.188 | 0.24 |
| 3 | 6537 | 0.209 | 0.000 | 0.005 | 0.001 | 0.367 | 0.368 | 0.790 | 0.001 | 0.338 | 0.24 |
| 4 | 1389 | 0.369 | 0.089 | 0.009 | 0.008 | 0.563 | 0.548 | 0.509 | 0.003 | 0.526 | 0.24 |
| 5 | 5465 | 0.570 | 0.291 | 0.025 | 0.012 | 0.806 | 0.749 | 0.668 | 0.007 | 0.746 | 0.24 |
| 6 | 3596 | 0.806 | 0.575 | 0.061 | 0.005 | 0.991 | 0.968 | 0.922 | 0.006 | 0.989 | 0.24 |
| 7 | 876 | 1.071 | 0.798 | 0.074 | 0.003 | 1.239 | 1.220 | 0.962 | 0.012 | 1.248 | 0.24 |
| 8 | 1314 | 1.356 | 0.916 | 0.085 | 0.001 | 1.506 | 1.504 | 0.994 | 0.007 | 1.524 | 0.24 |
| 9 | 217 | 1.655 | 0.966 | 0.082 | 0.000 | 1.802 | 1.804 | 0.997 | 0.011 | 1.816 | 0.24 |
| 10 | 517 | 1.962 | 0.986 | 0.082 | 0.000 | 2.105 | 2.109 | 0.999 | 0.010 | 2.120 | 0.24 |
| 11 | 392 | 2.271 | 0.994 | 0.079 | 0.000 | 2.407 | 2.415 | 1.000 | 0.012 | 2.427 | 0.24 |
| 12 | 423 | 2.579 | 0.997 | 0.077 | 0.000 | 2.707 | 2.717 | 1.000 | 0.012 | 2.732 | 0.24 |
| 13 | 242 | 2.882 | 0.999 | 0.077 | 0.000 | 3.002 | 3.011 | 1.000 | 0.010 | 3.030 | 0.24 |
| 14 | 136 | 3.176 | 0.999 | 0.075 | 0.000 | 3.285 | 3.294 | 1.000 | 0.011 | 3.320 | 0.24 |
| 15 | 126 | 3.461 | 1.000 | 0.073 | 0.000 | 3.557 | 3.565 | 1.000 | 0.011 | 3.598 | 0.24 |
| 16+ | 246 | 4.072 | 1.000 | 0.071 | 0.000 | 4.149 | 4.149 | 1.000 | 0.012 | 3.865 | 0.24 |

Age $\mathbf{0 , 1 , 2}$ over-written as follows:
2019 yc 2019 age 0 replaced by 2005-2016 LTGM (12 383);
2018 yc 2019 age 1 from SS3 survivor estimate at-age 1, 2019 * LTGM / SS3 estimate of age 0 in 2018;
2017 yc 2019 age 2 from SS3 survivor estimate at-age $2,2019 *$ LTGM / SS3 estimate of age 0 in 2017.

Table 30. Bss.27.4.bc, 7.a, d-h: Detailed short-term status quo forecast.


Table 31. Management options table. The F-at-age in 2015, when the French pelagic fishery was substantially reduced, was assumed as status quo for 2016 when the pelagic fishery was closed in spring, and assumed to continue in 2017. FMultipliers for 2017 are applied to both the commercial and recreational fishery. Note that the combined total commercial and recreational forecasted catch could be allocated in different ways.

| 2019 |  | Commercial fishery |  |  |  |  |  | Recreational fishery |  |  | Total fishery Total Fbar Total catch |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | Fmult | Fbar | Landings | Fmult | Fbar | Discards | Fmult | Fbar | Catch |  |  |
| 17540 | 10884 | 1.000 | 0.067 | 915 | 1.000 | 0.002 | 73 | 2.031 | 0.019 | 260 | 0.088 | 1248 |


| 2020 |  | Commercial fishery |  |  |  |  |  | Recreational fishery |  |  | Total fisheryTotal Fbar Total catch |  | 2021 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | Fmult | Fbar | Landings | Fmult | Fbar | Discards | Fmult | Fbar | Catch |  |  | Biomass | SSB |
| 17925 | 11413 | 1.901 | 0.127 | 1757 | 1.901 | 0.005 | 109 | 2.125 | 0.040 | 561 | 0.171 | 2427 | 16946 | 10475 |
|  |  | 1.901 | 0.127 | 1757 | 1.901 | 0.005 | 109 | 2.125 | 0.040 | 561 | 0.171 | 2427 | 16946 | 10475 |
|  |  | 1.470 | 0.098 | 1379 | 1.470 | 0.004 | 85 | 2.125 | 0.040 | 569 | 0.142 | 2033 | 17311 | 10794 |
|  |  | 0.000 | 0.000 | 0 | 0.000 | 0.000 | 0 | 0.317 | 0.006 | 91 | 0.006 | 91 | 19109 | 12371 |
|  |  | 2.049 | 0.136 | 1885 | 2.049 | 0.005 | 117 | 2.125 | 0.040 | 558 | 0.182 | 2560 | 16824 | 10367 |
|  |  | 3.100 | 0.206 | 2753 | 3.100 | 0.008 | 173 | 2.125 | 0.040 | 540 | 0.254 | 3466 | 15988 | 9638 |
|  |  | 2.189 | 0.146 | 2009 | 2.189 | 0.005 | 124 | 1.878 | 0.035 | 492 | 0.187 | 2625 | 16762 | 10313 |
|  |  | 1.000 | 0.067 | 963 | 1.000 | 0.002 | 59 | 1.000 | 0.019 | 275 | 0.088 | 1296 | 17991 | 11388 |
|  |  | 1.370 | 0.091 | 1290 | 1.370 | 0.003 | 79 | 2.125 | 0.040 | 571 | 0.135 | 1939 | 17397 | 10870 |
|  |  | 1.370 | 0.091 | 1290 | 1.370 | 0.003 | 79 | 2.125 | 0.040 | 571 | 0.135 | 1939 | 17397 | 10870 |
|  |  | 1.040 | 0.069 | 990 | 1.040 | 0.003 | 60 | 2.125 | 0.040 | 577 | 0.112 | 1627 | 17687 | 11123 |

Table 32. Annual average cpue bars Group 0 ( 1000 minutes trawling) and annual deviations from the time-series average per site. The sites are listed from north to south.

|  | annual LPUE (number of age 0 for 1000minutes of trawling |  |  |  |  |  |  |  | average annual deviation |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | area | 2005 | 2006 | 2007 | 2009 | 2010 | 2011 | average per area | 2005 | 2006 | 2007 | 2009 | 2010 | 2011 |
|  | seine aval |  | 4 |  |  | 133 | 15 | 51 |  | -91 |  |  | 161 | -70 |
| East Channel | Ome |  | 206 |  |  | 164 | 268 | 213 |  | -3 |  |  | -23 | 26 |
|  | Baie des Veys | 0 | 167 |  |  | 96 | 4 | 89 | -100 | 88 |  |  | 7 | -95 |
| West Channel | Mont St Michel |  | 567 |  |  | 836 | 252 | 551 |  | 3 |  |  | 52 | -54 |
| W | Morlaix |  |  | 664 | 182 | 535 | 456 | 459 |  |  | 45 | -60 | 16 | -1 |
|  | Laita |  |  | 0 | 2 | 278 | 17 | 74 |  |  | -100 | -98 | 275 | -78 |
| South Britanny | Blavet |  |  | 25 | 42 | 19 | 58 | 36 |  |  | -32 | 17 | -46 | 61 |
|  | Vilaine |  |  | 301 | 19 | 23 | 101 | 111 |  |  | 171 | -83 | -79 | -9 |
|  | Loire |  | 151 |  | 192 | 0 | 30 | 93 |  | 62 |  | 106 | -100 | -68 |
|  | Sevre Niortaise |  |  | 3772 | 2133 | 460 | 74 | 1610 |  |  | 134 | 32 | -71 | -95 |
|  | Charente |  |  |  | 28 | 14 | 6 | 16 |  |  |  | 76 | -12 | -65 |
| Bay of Biscay | Seudre | 0 |  |  | 127 | 0 | 11 | 35 | -100 |  |  | 268 | -100 | -68 |
|  | Gironde aval |  |  |  |  | 87 | 7 | 47 |  |  |  |  | 86 | -86 |
|  | Gironde | 3 |  |  | 72 |  |  | 38 | -91 |  |  | 91 |  |  |
|  | Adour aval | 4 | 22 |  | 12 | 0 | 0 | 8 | -45 | 191 |  | 54 | -100 | -100 |
|  |  |  |  |  |  |  |  | mean | -84 | 42 | 44 | 40 | 5 | -50 |
| SD >-20\% |  |  |  |  |  |  |  | SD | 26.2 | 96 | 112.6 | 108.1 | 109.6 | 49.8 |
| -20\%<SD>20\% |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SD >+20\% |  |  |  |  |  |  |  |  |  |  |  |  |  |  |



Figure 1. Bss.27.4.bc, 7.a, d-h. Trends in Official landings by country.


Figure 2. Bss.27.4.bc, 7.a, d-h. Trends in ICES estimates landings by gear (France-top and UK-bottom).


Figure 3. Bss.27.4.bc, 7.a, d-h. Length composition all French fleet combined from 2000 onwards.


Figure 4. Bss.27.4.bc, 7.a, d-h. Length composition of UKOTB_Nets fleets landings from 1985 onwards.

length comp data, retained, Lines

length comp data, retained, Lines


Figure 5. Length composition of UK Lines fleet landings from 1985 onwards.
length comp data, retained, UKMWT


Figure 6. Available length composition of UK Midwater pair trawl fleet landings.


Figure 7. Length frequency of recreational fishery removals for the 2012 reference year, derived from surveys in France, Netherlands and England. PRM are total released catch with post-release mortality of $5 \%$ applied. Right hand plot is the total removals used in the Stock Synthesis model to estimate selectivity.


Figure 8. Bss-47 stock: (a) Natural mortality values inferred from Gislason et al. (2010) and Lorenzen (1996); (b) Gislason M values rescaled to average $\mathrm{M}=\mathbf{0 . 1 5}$ or 0.24 at ages $\mathbf{1 0 - 2 0}$ ); (c) Lorenzen M values rescaled to average $\mathrm{M}=0.15$ or 0.24 at ages 10-20).


Figure 9. Location of Cefas Solent and Thames juvenile seabass surveys.

(b) Solent 1-gp index


Figure 10. Cefas Solent survey in autumn: (a) year and year-class effects in indices; (b) 1-gp index from 1996 onwards compared with a composite year-class index derived from the age 2-4 indices.


Figure 11. Left: stations fished during the Channel Groundfish Survey carried out annually by France. Right: distribution of total catches of seabass over the survey series.


Figure 12. Mean standardised time-series of (a) percentage of stations with seabass, and (b) swept-area abundance indices (millions of fish) from the Ifremer Channel Groundfish Survey.


Figure 13. Comparison in trends in commercial lpue index for French fleets for European seabass in ICES divisions 4bc and 7a,d-h (WGCSE 2019 Ipue corrected and updated in assessment)


Figure 14. Commercial lpue index for French fleet used in assessment with bootstrap intervals.


Figure 15. Left: Datasets used in the final seabass update assessment. Right: Catch series for the six fleets.



Figure 16a. Final seabass update assessment: Fitted length-based and age-based selectivity curves.


Figure 16b. Final seabass update assessment: Fitted time-series of length-based and age-based selectivity and retention curves for fleets with blocks.


Figure 17. Final seabass update assessment: fit to UK trawl and net fishery-length composition data for the retained and discarded catch components.


Figure 18. Final seabass update assessment: fit to UK lines length-composition data for the retained catch components.


Figure 19. Final seabass update assessment: fit to UK midwater trawl fishery length-composition data for the retained catch components.


Figure 20. Final seabass update assessment: fit to French fishery length-composition data for the retained and discarded catch components.


Figure 21. Final seabass update assessment: Fit to recreational length-compositions data.


Pearson residuals, whole catch, CGFS1 (max=10.53)


Figure 22. Final seabass update assessment: Fit to Channel groundfish survey length-compositions.

## length comps, whole catch, aggregated across time by fleet



Figure 23. Final seabass update assessment: Fit to the commercial fisheries and Channel groundfish survey length compositions, aggregated across time for the retained and discarded catch components.


Figure 24. Final seabass update assessment: Fit to age composition data for the combined UK otter trawl and nets fleets.


Figure 25. Final seabass update assessment: Fit to age composition data for the combined UK lines fleet.


Pearson residuals, retained, UKMWT (max=6.44)


Figure 26. Final seabass update assessment: Fit to age composition data for the UK midwater trawl fleet.


Pearson residuals, retained, French (max=2.63)


Figure 27. Final seabass update assessment: Fit to age composition data for the combined French fleets.


Pearson residuals, whole catch, AutBass (max=3.78)


Figure 28. Final seabass update assessment: Fit to age composition data for the Solent Autumn bass survey.



Figure 29. Final seabass update assessment: Fit to UK fleets age compositions, aggregated across time.


Figure 30. Final seabass update assessment: Fit to Solent Autumn bass survey total abundance index, accounting for age and length-based selectivity.


Figure 31. Final seabass update assessment: Fit to Channel groundfish survey total abundance index, accounting for length-based selectivity.


Figure 32. Final seabass update assessment: Fit to the French landings per unit of effort commercial index, accounting for length-based selectivity.


Figure 33. Final seabass update assessment: Top: time-series of log-recruit deviations (deviations for 1965-1984 precede the period of input catch data). Below: stock-recruit scatter (model is fitted assuming Beverton-Holt stock-recruit model and steepness $=0.999$ ).


Figure 34. Retrospective analysis of stock trends from final update assessment, based on Stock Synthesis run final year set to 2015 and peeling back five years (for the final run, terminal F is for $\mathbf{2 0 1 8}$ and SSB and total biomass terminate in 2019).


Figure 35. Stock trends from final update assessment, based on Stock Synthesis run final year set at 2018 to give 2019 numbers and biomass and 2018 F. Recruitment in 2017-2019 is the geometric mean 2005-2016. Recruitment, F and SSB are shown with $95 \%$ confidence intervals.


Figure 36. Comparison between LPUE and recreational catch vector from this year's final update assessment and the 2018 WGCSE assessment.


Figure 37. Comparison between stock trends from this year's final update assessment and the 2018 WGCSE assessment.

# 30 Seabass (Dicentrarchus labrax) in divisions 6.a,7.b and 7.j (West of Scotland, West of Ireland, eastern part of southwest of Ireland) 

## Type of assessment

There is no assessment for this stock component.

ICES advice applicable to 2018, 2019 \& 2020
"Based on ICES approach to data-limited stocks, ICES advises that when the precautionary approach is applied, commercial landings should be no more than 4 tonnes in each of the years 2018, 2019 and 2020. ICES cannot quantify total catches. No information on discards is available, therefore it is not possible to provide commercial catch advice. Also, recreational catches cannot be quantified. Therefore, total catches cannot be calculated.

## ICES advice applicable to 2016

"The revised landings data do not change the perception of the stock but result in a revision of the advised landings. Therefore, ICES advises based on the data-limited stocks approach, but cannot quantify the resulting catches. The implied commercial landings should be no more than 5 tonnes.

Currently, there is no TAC for this species in this area, and it is not clear whether this should constitute a separate management unit. ICES does not necessarily advocate the introduction of a TAC for sea bass in this area."

### 30.1 General

## Stock description and management units

At IBP-NEW (2012a), it was agreed that sea bass in the North Sea (4.b\&c) and in the Irish Sea, Channel and Celtic Sea (7.a,d,e,f,g\&h) would be treated as a functional stock unit as there is no clear basis from fishery data, tagging and genetics studies to subdivide the populations in the Irish Sea, Celtic Sea, Channel and North Sea into independent stock units. It was proposed based on previous ICES bass study group reports to allocate sea bass in 6.a, 7.b and 7.j to a separate stock, although it is recognised that sea bass in Irish coastal waters of 7.g and 7.a are likely to be from the same stock as in 7.j. As there are negligible commercial fishery catches of sea bass in Irish coastal waters due to the moratorium on commercial fishing for bass by Irish vessels, the splitting of the stock between 7.g and is not likely to have any impact on the bass assessment in $4 . \mathrm{b}, \mathrm{c}$ and $7 . \mathrm{a}, \mathrm{d}-\mathrm{h}$. Supporting information can be found in the IBP-NEW (ICES, 2012a) report.

Management applicable to 2016, 2017 and 2018
Sea bass are not subject to EU TACs and quotas. A moratorium on commercial fishing for sea bass has been in place for Irish vessels fishing in areas 6 and 7 since 1990, and a minimum landing size of 40 cm applies to Irish fisheries. The official minimum landing size for non-Irish vessels is 36 cm (EC regulation 850/98). In addition, a variety of national restrictions on commercial sea bass fishing are also in place for non-Irish commercial vessels, including licensing, individual
landings limitations, larger MLS and seasonal/ area closures. Recreational fishing for sea bass in Ireland is prohibited from 15 May to 15 June, and a bag limit of two fish per 24 hours is in place.

Previous advice from ICES, showing a rapid decline in sea bass biomass in the North Sea, Channel, Celtic Sea and Irish Sea caused by poor recruitment and over-fishing, has resulted in the European Commission working with Member States to identify more effective control measures to reduce fishing mortality towards FMSY. It shall be prohibited for Union fishing vessels to fish for sea bass in ICES divisions 7.b, 7.c, 7.j and 7.k, as well as in the waters of ICES divisions 7.a and 7.g that are more than 12 nautical miles from the baseline under the sovereignty of the UK. It shall be prohibited for Union fishing vessels to retain on board, tranship, relocate or land sea bass caught in that area. Depending on the true stock structure of sea bass in areas 6 and 7, very restrictive measures introduced in 2016, 2017 and 2018 may have some effect on sea bass in 6.a, 7.b and 7.j.

## Fishery in 2018

Landings data used by the WG are given in Table 32.2.1. Due to the Irish sea bass moratorium, official landings reports are by other countries, historically mainly by France, although the landings are less than 10 tonnes per year and only 2 tonnes or less since 2012. No landings have been recorded for 2017 and only 1 tonne in 2018 (source: official landings).

### 30.2 Data

## Commercial landings data

Landings data are given in Table 32.2.1. No other data for sea bass in this area were provided to WGCSE.

## Commercial discards

No estimates of sea bass discards are available.

## Recreational catches

Recreational marine fishery surveys in Europe are still at an early stage in development and are described by the ICES Working Group on Recreational Fishery Surveys (ICES, 2012b). A survey was conducted in Ireland in 2010 and 2011 (O'Reilly and Roche, 2012). Domestic shore bass anglers are estimated at 11600 individuals and these anglers harvested and estimates of 30 t and 44 t of bass in 2010 and 2011. The 2010 estimate was considered to be more robust. In addition between $75 \%$ and $80 \%$ of bass caught were returned to the water. The survey doesn't disaggregate the angling catch estimates by ICES division.

The IBP-NEW meeting report (ICES, 2012a) includes some data supplied by a stakeholder on trends in recreational catch rates from an angling club on the southern Irish coast, as well as age compositions of sea bass caught by anglers, which may be applicable also to trends in 7.j.

## Biological data

Data on growth and maturity for this stock component were not reviewed by WGCSE.

## Survey data

No survey data were available to WGCSE for this stock.

## Other relevant data

None.

### 30.3 Historical stock development

No information is available for this stock area.

### 30.4 Management plans

There are no existing management plans for European sea bass.

### 30.5 Management considerations

Sea bass grow slowly, do not mature until 4-7 years of age, and have been recorded at up to 28 years of age. Juvenile bass up to three years of age, occupy nursery areas in estuaries whilst adults undertake seasonal migrations from inshore habitats to offshore spawning sites. It is not known to what extent adults from the stock in $7 . b, j$ and 6. a are caught by pelagic trawlers targeting mature sea bass on spawning sites in divisions 7.e-h. After spawning, sea bass tend to return to the same coastal sites each year. The combination of slow growth, late maturity, spawning aggregation and strong site fidelity, increase the vulnerability of sea bass to overexploitation and localized depletion.

ICES advice sheets for sea bass in the Northeast Atlantic have previously recommended that "implementation of 'input' controls (preferably through technical measures aimed at protecting juvenile fish, in conjunction with entry limitations into the offshore fishery in particular) should be promoted (ICES, 2004)" and that "Any consideration of catch limitation (output control) would need to take into account that sea bass are a bycatch in mixed fisheries to a various extent, depending on gear and country; this incites discarding and should be avoided".

Management of sea bass fisheries needs to take into account the distinctive characteristics and economic value of the different fisheries. Sea bass is of high social and economic value to sea angling in Ireland which contributes substantially to local economies.

The current stock structure assumptions are pragmatic, and need further evaluation. Further studies are needed to determine if the sea bass in Irish coastal waters are indeed functionally separate, or if they also mix with the other stock during spawning time and contribute to commercial catches on the offshore spawning grounds.

As bass is, at present, a non-TAC species, there is potential for displacement of fishing effort by non-Irish fleets from other species with limiting quotas.

### 30.6 Data needs

Time-series of relative abundance indices need to be developed throughout the range of the stock, for both the adult and pre-recruit components of the stock.

There is a need to develop a time-series of recreational fishery catch, effort, and catch composition.

Catch locations and composition of significant commercial landings should be monitored to help establish the stock affiliation.

Further studies using tagging, genetics, and other stock and individual markers are needed to more accurately define stock boundaries suitable for assessment and management purposes. A tagging programme has been undertaken by the Marine Institute of Ireland to investigate the distribution of the European sea bass in Irish waters. This project is being carried out in conjunction with the Beaufort Scientific Group and University College Cork. No results were available for WGCSE 2019.

Studies are needed to document the survival of recreationally caught and released sea bass. IBPNEW (ICES, 2012a) noted that a range of studies on striped bass in the USA indicated hooking mortalities of around $20 \%$ on average, although a lower value of around $9 \%$ from one specific study is currently considered most appropriate for inclusion in the assessments.

### 30.7 References

ICES. 2012a. Report of the Inter-Benchmark Protocol on New Species (Turbot and Sea bass; IBPNew 2012). ICES CM 2012/ACOM:45.

ICES. 2012b. Report of the Working Group on Recreational Fisheries Surveys (WGRFS). ICES CM 2012/ACOM:23. 55 pp.

O'Reilly, S. and Roche, W. 2012. Pilot study to estimate recreational angling landings of bass in Ireland. Inland Fisheries Ireland report IFI/2012/1-4099. http://www.miextranet.ie/fss/sites/DCMAP/Annual\ Report/Annex 2 DCF Bass Landings 2010 11.pdf.

Table 32.2.1. European sea bass in Divisions 6.a, 7.b and 7.j. Official landings: all countries (predominantly France).

| Year | Official landings |
| :---: | :---: |
| 2000 | 1 |
| 2001 | 4 |
| 2002 | 4 |
| 2003 | 2 |
| 2004 | 8 |
| 2005 | 4 |
| 2006 | 2 |
| 2007 | 5 |
| 2008 | 5 |
| 2009 | 4 |
| 2010 | 9 |
| 2011 | 7 |
| 2012 | 1 |
| 2013 | 0 |
| 2014 | 2 |
| 2015 | 0.8 |
| 2016 | 0.1 |
| 2017 | 0 |
| 2018* | 1 |

*Preliminary. Source Official landings.

## 31 Sole in Division 7.a (Irish Sea)

## Type of assessment in 2019

This assessment is an update assessment.

## ICES advice applicable to 2019

In the advice for 2019, the stock status was presented as follows:

"ICES advises that when the MSY approach is applied, catches in 2019 should be no more than 414 tonnes".

## Comments made by the audit of last year's assessment

No major deficiencies for the sole assessment in the Irish Sea were reported.

### 31.1 General

## Stock description and management units

The sole fisheries in the Irish Sea are managed by TAC (see text tables below) and technical measures, with the assessment area corresponding to the stock area. Technical measures in force are minimum mesh sizes and minimum landing size ( 24 cm ). In addition beam trawlers, fishing with mesh sizes equal to or greater than 80 mm , are obliged to have 180 mm mesh sizes in the entire upper half of the anterior part of their net. More details can be found in Council Regulation (EC) $\mathrm{N}^{\circ} 254 / 2002$ and the Stock Annex.

Since 2000, a spawning closure for cod has been in force. The first year of the regulation the closure covered the western and eastern Irish Sea. Since then, closure has been mainly in the western part whereas the sole fishery takes place mainly in the eastern part of the Irish Sea (Liverpool Bay and Cardigan Bay). No direct impact on the sole stock is expected from this closure.

For 2009 Council Regulation (EC) $\mathrm{N}^{\circ} 43 / 2009$ allocates different amounts of $\mathrm{kW}^{*}$ days by Member State and area to different effort groups of vessels depending on gear and mesh size. The areas are Kattegat, part of 3.a not covered by Skaggerak and Kattegat, ICES zone 4, EC waters of ICES zone 2.a, ICES zone 7.d, ICES zone 7.a, ICES zone 6.a and EC waters of ICES zone 5.b. The grouping of fishing gear concerned are: bottom trawls, Danish seines and similar gear, excluding beam trawls of mesh size: TR1 ( $\geq 100 \mathrm{~mm}$ )-TR2 ( $\geq 70$ and $<100 \mathrm{~mm}$ )-TR3 ( $\geq 16$ and $<32 \mathrm{~mm}$ ); beam trawl of mesh size: BT1 ( $\geq 120 \mathrm{~mm}$ )-BT2 ( $\geq 80$ and $<120 \mathrm{~mm}$ ); gillnets excluding trammelnets: GN1; trammelnets: GT1 and Longlines: LL1.

For 2010-2016, Council Regulation (EC) N ${ }^{\circ} 53 / 2010$, Council Regulation (EC) N ${ }^{\circ} 57 / 2011$, Council Regulation (EC) $\mathrm{N}^{\circ} 43 / 2012$, Council Regulation (EC) $\mathrm{N}^{\circ} 40 / 2013$, Council Regulation (EC) $\mathrm{N}^{\circ} 43 / 2014$, Council Regulation (EC) $\mathrm{N}^{\circ} 2015 / 104$ and Council Regulation (EC) N ${ }^{\circ} 2016 / 72$ were updates of the Council Regulation (EC) $N^{\circ} 43 / 2009$ with new allocations, based on the same effort groups of vessels and areas as stipulated in Council Regulation (EC) N²43/2009.

Management applicable to 2018 and 2019

TAC 2018

| Species:Common sole <br> Solea solea |  | Zone: | $\begin{aligned} & \text { 7a } \\ & \text { (SOL/07A.) } \end{aligned}$ |
| :---: | :---: | :---: | :---: |
| Belgium | $10{ }^{1}$ ) |  |  |
| France | $0{ }^{(1)}$ |  |  |
| Ireland | $17{ }^{(1)}$ |  |  |
| The Netherlands | $3{ }^{(1)}$ |  |  |
| United Kingdom | $10{ }^{(1)}$ |  |  |
| Union | $40{ }^{(1)}$ |  |  |
| TAC | $40{ }^{(1)}$ |  | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 shall not apply <br> Article 4 of Regulation (EC) No 847/96 shall not apply |

${ }^{(1)}$ Exclusively for by-catches. No directed fisheries are permitted under this quota.

TAC 2019

| Species: $\quad \begin{aligned} & \text { Common sole } \\ & \text { Solea solea }\end{aligned}$ |  | Zone: | $\begin{aligned} & 7 \mathrm{a} \\ & \text { (SOL/07A.) } \end{aligned}$ |
| :---: | :---: | :---: | :---: |
| Belgium | 192 |  |  |
| France | 2 |  |  |
| Ireland | 74 |  |  |
| The Netherlands | 60 |  |  |
| United Kingdom | 86 |  |  |
| Union | 414 |  |  |
| TAC | 414 |  | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 shall not apply <br> Article 4 of Regulation (EC) No 847/96 shall not apply |

## Fishery in 2018

A full description of the fishery is provided in the Stock Annex, Section A2.
An overview of the landings data provided and used by the Working Group (WG) is shown in Table 29.1. The landings reached a level of 2808 t in the mid-1980s due to good recruitments in 1982-1984, but then subsequently dropped to a lowest of 818 t in 2000. After a small increase to 1090 t in the beginning of the 2000s, the landings have fallen to under 350 t in 2008-2012. From 2013 onwards the landings continued to decrease as they dropped to under 150 t .

The WG estimated the total international landings at 36 t in 2018, of which Belgium landed $40 \%$ (14.3 t), Ireland $46 \%(16.5 \mathrm{t}), 9 \%(3.2 \mathrm{t}$ ) by the UK (England \& Wales) and the remainder by Northern Ireland, the Netherlands, Scotland, Isle of Man and France. These landing-figures are among the lowest in the time-series, corresponding to an international uptake of $90 \%$ of the agreed TAC in 2018 (40 t) and last year's forecast.

The WG estimate of the 2017 landings was not revised.
In $2018,66 \%$ of the landings were taken by beam trawls, $30 \%$ by otter trawls, $4 \%$ by other gears.

### 31.2 Data

## Landings

Quarterly age compositions for 2018 were available from the countries that take the major part of the international landings ( $95 \%$ ) (Belgium, UK(E\&W) and Ireland). The raw age data were combined for the three countries without weighting. The combined ALK was applied to the raised length distribution of the national catches to obtain a combined age distribution. This distribution was applied to the landings from Northern Ireland, Scotland, Isle of Man and France to obtain the catch numbers-at-age for 2018 (Table 29.2, Figure 29.1). The standardised catch pro-portion-at-age is presented in Figure 29.2. Annual length distributions of the three major countries involved are given in Table 29.3. Because of the substantial reduction of the TAC in the last four years, sampling levels are also substantially reduced.
Catch weights-at-age for 2018 were taken from the combined age-weight key (Table 29.4).
Stock weights-at-age for 2000-2018 were derived from the mean catch weights by cohort interpolation to the first of January (Rivard weight calculator) (Table 29.5).

Further details on raising methods are given in the stock annex.
As last year, the combined age data (calculated outside InterCatch) as well as the landings from Northern Ireland, Scotland, Isle of Man and France were uploaded to InterCatch. It should be noted that the international age distribution is uploaded as "BE" as no international country code is available in InterCatch at present.

## Discards

The available discard information (Table 29.6) suggests that discarding is not a major problem in the Irish Sea sole fishery. Belgian beam trawl length distributions of retained and discarded catches of sole for 2018 (Figure 29.3a) indicate that predominantly 3 and 4-year old fish are discarded. Observer information from UK and Irish otter trawl fleets also suggest low discard rates. The working group decided not to include discards in the assessment at this stage due to the scarcity of the data but will monitor the situation in the future.

As an attempt, estimating an overall discard rate for the stock, individual discard estimates for 2016-2018 from the main métiers and countries were averaged to obtain an overall discard rate (Table 29.6b). The percent of the métiers with discard information covering the total international landings is $65 \%, 63 \%$ and $66 \%$ for 2016, 2017 and 2018 respectively. Assuming that discard rates do not change from the average of the last three years (2016-2018) and a fixed proportion of discards survive, a discard rate of around $3.5 \%$ (of the catch) could be assumed for this stock at the moment.

## Biological

Natural mortality, maturity and proportions of natural mortality and fishing mortality before spawning were set as in previous years, details of which can be found in the Stock Annex Section B2.

## Surveys

Lpue and effort series were available from the UK(E\&W) September beam trawl survey (UK(E\&W)-BTS-Q3) (1988-2018) and the UK(E\&W) March beam trawl survey (UK(E\&W)-BTS-

Q1) (1993-1998) (Table 29.7b and Figure 29.4c). From 2006 until 2010 the two UK beam trawl surveys have been used as tuning indices in the Irish Sea sole assessments. Following the outcome of WKFLAT 2011, the March survey (UK(E\&W)-BTS-Q1) was omitted from the following assessments. The lpue from the UK(E\&W)-BTS-Q3 has fluctuated since the beginning of the timeseries (1988) between 90 and $200 \mathrm{~kg} / 100 \mathrm{Km}$ fished. Since 2000, it has dropped gradually to the lowest value in 2012 ( $26.47 \mathrm{~kg} / 100 \mathrm{Km}$ fished). Thereafter, it slowly increased to $78.51 \mathrm{~kg} / 100 \mathrm{Km}$ fished in 2018.

Detailed information on the survey protocols and area coverage can be found in the Stock Annex.

## Commercial Ipue

Trends in lpue and effort are given in Table 29.7 and Figure 29.4-29.5.
Commercial lpue and effort data were available for Belgian beam trawlers, UK(E\&W) beam and otter trawlers and Irish otter and beam trawlers. It should be noted that the most recent lpue values of the UK(E\&W) beam trawlers (2013-2018) and the UK(E\&W) otter trawlers (2014-2018) are not available as the effort values for those years are missing. In 2013, the UK administration switched to the EU electronic logbook system. Therefore, a lot of the reported effort is missing and the 2013 value cannot be used as an absolute number. Details of the 2013 UK beam trawl were unavailable due to reduced numbers of trips reporting this gear specific effort information via the newly introduced e-logbook system. The otter trawl fleet effort reporting was unaffected by this as these vessels were not reporting their landings via this method in 2013. However, from 2014 onwards, both the UK beam trawl and otter trawl effort values are unavailable because of the reporting issues.

Effort from both Belgian and UK commercial beam trawl fleets increased from the early seventies until the beginning of the nineties. Since then UK beam trawl effort has shown a continuing declining trend. Inspection of an alternate effort indicator (days fished) suggests that the declining trend continues in the period 2013-2018. In contrast, the Belgian beam trawl effort has shown a fluctuating pattern. After the decline in the early nineties, it reached its highest level in 2002 and decreased again afterwards. For the period 2008-2012, it remained stable at a very low level but in 2013, it continued to decrease and in 2016 it dropped to the lowest level in the time-series. In 2017-2018, there is a slight increase. The effort of the Irish beam trawlers shows a slow decline since 2004 and reached the lowest level in the time-series in 2013. Since 2014, the Irish beam trawl effort has slightly increased. In 2008 all beam trawl fleets showed a substantial reduction in effort compared to 2007.

The effort from the UK otter trawlers remained stable until the beginning of the nineties. Since then, the UK otter trawl effort has continuously declined and is at the lowest level in 2013. As, in 2015 and 2016 all otter trawl vessels active in the Irish Sea were under 12 m , no effort (days fished) was recorded. In 2017, the otter trawl effort (days fished) is similar to the level observed in the period 2009-2010 and in 2018 it further increased. The Irish otter trawlers have shown a striking reduction in effort since 2000, followed by a slight increase in the period 2010-2012. In 2017 and 2018, the Irish otter trawl effort fell back to the lowest observed levels in the time-series.

Nearly all effort time-series show a substantial decrease in the last six years, in line with the substantial reductions of the TAC.

Lpue for both UK and Belgian beam trawlers was at a high level in the late seventies and early eighties but since early 2000s, lpue for these fleets has fluctuated at a lower level. In the period 2007-2009 there has been a small increase in the UK beam trawl lpue. However, in 2012 the lpue has dropped to a remarkable low level in the time-series ( $4.3 \mathrm{~kg} / \mathrm{hour}$ fished). An update for 2013-2017 was not available. However, the alternate lpue indicator (kg/days fished) suggests that the UK beam trawl lpue increased in 2015. For 2016-2018 no catches of sole and/or no effort
were recorded therefore the lpue is zero. The Belgian beam trawlers hold on to a higher lpue value ( $18-20 \mathrm{~kg} / \mathrm{hour}$ fished) for the period 2008-2012. However, in 2013 the lpue decreased ( $12.7 \mathrm{~kg} / \mathrm{hour}$ fished) and in 2017 it dropped to the lowest level in the time-series ( $3.6 \mathrm{~kg} / \mathrm{hour}$ fished). In 2018, there's a slight increase to $5.4 \mathrm{~kg} /$ hour fished. The Irish beam trawl lpue shows a gradually diminishing trend over the whole time-series. After the slight increase in 2013, it fell back to a record low level in 2016-2018.

The UK otter trawl lpue remained stable until the beginning of the 2000s but is at the record low level in 2012. The alternative lpue indicator (kg/days fished) suggests that the declining trend continues after 2012. In 2018 there is a slight increase to the level observed in the period 20112012 ( $12.6 \mathrm{~kg} /$ days fished). In 2012-2016, the lpue of Irish otter trawlers is fluctuating at a lower level. In 2017-2018 a higher value was recorded.

## Historical stock development

In 2010, the Irish Sea sole assessment was based on XSA with two survey tuning indices (UK(E\&W)-BTS-Q3 and UK(E\&W)-BTS-Q1 (Table 29.8). The UK(E\&W)-BTS-Q1 indices only provide information for the years 1993 up to 1999 and therefore no longer contribute to the final survivor estimates. At WKFLAT 2011, the exclusion of the UK(E\&W)-BTS-Q1 from the assessment was investigated and it was found that there was little effect on the catchability residuals and that the retrospective pattern was slightly improved. WKFLAT 2011 therefore decided to omit this survey from the assessment.

### 31.3 Stock assessment

## Data screening

The age range for the analysis was $2-8+$.
The screening of the tuning indices (UK(E\&W)-BTS-Q3) showed good cohort tracking (Figure 29.6) and consistency between ages for year-class strength (Figure 29.7).

## Final update assessment

The model settings for the final assessment are summarized below:

| Assmnt Year | :2010 | :2011-2019 |
| :---: | :---: | :---: |
| Assmnt Model | : XSA | :XSA |
| Fleets | : | : |
| Bel Beam Trwl | : omitted | :omitted |
| UK Trawl | : omitted | :omitted |
| UK Sept BTS | :1988-2009 2-7 | :1988-2018 2-7 |
| UK Mar BTS | :1993-1999 2-7 | :omitted |
| Time Ser. Wts | : linear 20 yrs | :no taper weighting |
| Power Model | : none | :none |
| Q plateau | : 7 | :4 |
| Shk se | :1.5 | :1.5 |
| Shk age-yr | : 5 yrs 3 ages | : 5 yrs 3 ages |
| Pop Shk se | : 0.3 | : 0.3 |
| Prior Wting | : none | : none |
| Plusgroup | : 8 | : 8 |
| Fbar | : 4-7 | : 4-7 |

The final XSA output is given in Table 29.9 (diagnostics), Table 29.10 (fishing mortalities) and Table 29.11 (stock numbers). Log catchability residuals for the final assessment are given in Figure 29.8. A summary of the XSA results is given in Table 29.12 and trends in yield, fishing mortality, recruitment and spawning stock biomass are shown in Figure 29.9. Retrospective patterns for the final run are shown in Figure 29.10.

Adding the 2018 data to the time-series did not cause any additional anomalies compared to last year. The log catchability residual pattern showed no trends apart from the year effect in 2016. The positive residuals (higher estimates from the UK(E\&W)-BTS-Q3 fleet compared to the VPA estimates) in 2016 are likely due to the fact that de age composition in the catch is flattened.

The survivor estimates and fishing mortality estimates are almost entirely determined by the UK(E\&W)-BTS-Q3 survey as it gets a high weighting ( $>96 \%$ ) at all ages.

A Mohn's rho analysis was conducted based on the XSA stock assessment results, i.e. the last data year (2018) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (ages 4-7) | recruitment |
| :--- | :---: | :---: | :---: |
| Mohn's rho value | 0.042 | -0.04 | -0.025 |

The Mohn's rho values for this assessment are very low and well below the threshold of $20 \%$ imposed by ICES for 2019 assessments, i.e. the current assessment indicates a high consistency.

## Comparison with previous assessments

A comparison of the estimates of this year's assessment with last year's is given in Figure 29.11.
Trends in fishing mortality, SSB and recruitment are very similar. In last year's assessment, F and SSB for 2017 were estimated to be 0.0188 and 1941 t respectively; this year's estimates for 2017 are 0.0193 and 1891 t , an upward revision of $3 \%$ for F and a downward revision of $3 \%$ for SSB. The estimated recruitment by XSA in 2017 (1362 thousand fish) was revised upward by $14 \%$ in this year's assessment (1553 thousand fish).

## State of the stock

Estimated trends of Irish Sea sole landings, SSB, fishing mortality and recruitment are presented in Table 29.12 and Figure 29.9. Since the late eighties the landings of Irish Sea sole have been declining to the lowest level of the time-series ( 34 t ) in 2017. SSB has been at a higher level until the late eighties. Since then SSB has been fluctuating between $B_{p a}$ and $B_{l i m}$ and since 2004, it dropped below Blim. After the record low value in 2014 ( 883 t ), SSB gradually increased again to 2627 t in 2018. High fishing mortalities were observed during the late eighties until the midnineties. Thereafter fishing mortality declined to a level fluctuating just above Flim. From 2013 onwards, fishing mortality has dropped under the level of $\mathrm{F}_{\mathrm{p}} / \mathrm{F}_{\mathrm{ms}}$. In 2018 the lowest level of the time-series was recorded (0.014). The decline in $F$ is supported by a substantial reduction of the TAC in the most recent years. Since 2001 recruitment has been well below the mean (5538 thousand fish) and the 2011 recruitment (year class 2009) is estimated to be the lowest in the time-series ( 638 thousand fish). The 2016 recruitment ( 4422 thousand fish, year class 2014) is estimated to be seven times higher than the record low recruitment in 2011. The lower 2017 recruitment ( 1553 thousand fish) is followed by a higher incoming recruitment (year class 2016), estimated to be 3670 thousand fish.

### 31.4 Short-term projections

## Estimating year-class abundance

The 2016 year class is now estimated at 3670 thousand fish at age 2, which is $162 \%$ higher than the short-term GM (2008-2016 (1402 thousand fish) used in the last year's forecast. The age 2 estimates are almost solely coming from the UK(E\&W)-BTS-Q3. From 2010 to 2014, the $\mathrm{UK}(\mathrm{E} \& \mathrm{~W})$-BTS-Q3 abundance for age 2 fluctuated around the level of the lowest abundance in 2011 (0.29). In 2016 the UK(E\&W)-BTS-Q3 abundance for age 2 increased to the higher level of early 2000 (2.97). In 2017, there is a decrease (0.8), followed by an increase in 2018 (2.18).

Given the consecutive low recruitments in recent years, the WG decided to assume the shortterm GM for the 2017 year class instead of the long-term GM (1970-2016, 4087 thousand fish). The short-term GM (2009-2017, 1339 thousand fish) recruitment was also assumed for the 2018 and subsequent year classes.

The working group estimates of year-class strength used for prediction can be summarised as follows:

| Year Class | XSA | GM 70-16 | GM 09-17 |
| :--- | :---: | :---: | :---: | :---: |
| 2016 (age 3 in 2019) | 3318 | 3595 | - |
| 2017 (age 2 in 2019) | - | 4087 | 1339 |
| $2018 \& 2019$ (recruits) | - | 4087 | 1339 |

Fishing mortality was calculated as the mean of 2016-2018. Catch and stock weights-at-age were also averages for the years 2016-2018. Population numbers at the start of 2019 for ages 3 and older, were taken from the XSA output.

In line with last year's forecast, the working group agreed to use a TAC constraint ( 400 t landings) for the intermediate year (2019). Because of the restricted fishing opportunities by the main countries fishing for Irish Sea sole, it seemed reasonable that the landings in 2019 would be in line with the agreed TAC of 414 t catch and 400 t landings.
The input for the short-term catch predictions and sensitivity analysis is given in Table 29.13, the short-term management option table is given in Table 29.14 and a detailed output is presented in Table 29.15.

Assuming a TAC constraint for 2019 of $414 \mathrm{t}(400 \mathrm{t}$ landings), implies a fishing mortality in 2019 of 0.133 . The assumed landings using a status quo fishing mortality in 2020 is 73 t . This results in a SSB of 3218 t in 2020 and 3481 t in 2021. The proportional contributions of recent year classes to the predicted landings and SSB are given in Figure 29.12. The assumed short-term GM recruitment accounts for about $5 \%$ of the landings in 2020 and about $15 \%$ of the 2021 SSB.

### 31.5 MSY explorations

Investigations for possible Fmsy candidates for this stock were carried out at WGCSE 2010. ACOM adopted an FmsY value of 0.16, based on stochastic simulations using a Ricker model (PLOTMSY program). $\mathrm{B}_{\text {trigger }}$ was set to the $\mathrm{B}_{\mathrm{pa}}$ value of 3100 t .
Exploratory analysis investigating possible revisions of MSY estimates were conducted at WGCSE 2014 with a recent version of PLOTMSY (Cefas, 2014). The simulations indicated the use of equally weighting for the stock-recruitment relationships and the resulting FmsY value was in line with the FmsY of 0.16 used at that moment for this stock.

In response to the EC long-term management plans for western EU waters (ICES subareas 5 to 10), ICES WKMSYREF4 (October 2015, Brest (France)) used long-term stochastic simulations (Eqsim) to estimate Fmš and appropriate ranges. The methodology used for stocks with agebased assessments follows the approaches developed in ICES WKMSYREF2 (ICES, 2014b) and WKMSYREF3 (ICES, 2014c) and is documented in the report of WKMSYREF4 (ICES, 2016c). Estimates of reference points Blim, $\mathrm{B}_{\mathrm{pa}}$, Flim and $\mathrm{F}_{\mathrm{pa}}$ were provided, and the Fmsy ranges [Flower, $\mathrm{F}_{\text {upper] }}$ ] deliver no more than $5 \%$ reduction in long-term yield compared with MSY.
The sole 7.a stock is at a low level and mean recruitment has been seen to be reduced at current biomass, simulations were conducted with S-R function (Beverton-Holt and Ricker models) that followed the mean of the recruitment data, giving some reduction in recruitment at Blim. The revised MSY reference points are less restrictive (Fmš $=0.20$ instead of 0.16 and MSY Btrigger $=3500 \mathrm{t}$ instead of 3100 t ).

In order to be consistent with the ICES precautionary approach, Fupper is capped, so that the probability of SSB $<$ Blim is no more than $5 \%$. Two approaches have been used to derive the values of
the cap on $F_{\text {upper }}$. One conforms to the ICES MSY advice rule (AR), and requires reducing F linearly towards zero when SSB is below MSY $B_{\text {trigger. }}$. The second uses a constant $F$ without an advice rule; i.e. no reduction in F with SSB less than MSY Btrigger. Although the first often provides a wider FMSY range, it requires the ICES MSY advice rule to be used (ICES, 2016d).

| Stock code | MSY Flower | F $_{\text {MSY }}$ | MSY Fupper with AR | MSY Fupper with no AR |
| :--- | :---: | :---: | :---: | :---: |
| Sol.7.a | 0.16 | 0.20 | 0.24 | 0.22 |

### 31.6 Biological reference points

## Precautionary approach reference points

The Working Group's current approach to reference points is outlined in Section 33.5. Current biological reference points are given in the text table below:

| Reference points | ACFM 2007 onwards | 2016 onwards |
| :---: | :---: | :---: |
| $\mathrm{F}_{\text {MSY }}$ | 0.16 (PLOTMSY, WG2010) | 0.20 (Eqsim, WKMSYREF 4) |
| $\mathrm{F}_{\text {lim }}$ | 0.4 (based on $\mathrm{F}_{\text {loss }}$ ) | 0.29 (based on simulated recruitment to give median biomass $=B_{\text {lim }}$ ) |
| $\mathrm{F}_{\text {PA }}$ | 0.3 (high probability of avoiding $\mathrm{F}_{\text {lim }}$ ) | 0.21 ( $\mathrm{Flim}^{* 1.4 \text { ) }}$ |
| $\mathrm{Blim}^{\text {l }}$ | 2200 ( ( $\mathrm{Bloss}^{\text {estimated in }}$ 2007) | 2500 t (lowest value with above average recruitment) |
| $B_{\text {PA }}$ | $3100 \mathrm{t}\left(\mathrm{B}_{\mathrm{pa}} \sim \mathrm{B}_{\mathrm{lim}} * 1.4\right)$ | 3500 t ( $\mathrm{Blim}^{*} 1.4$ ) |
| $\mathrm{B}_{\text {trigger }}$ | $B_{\text {PA }}$ | 3500 t |

### 31.7 Management Plans

No management plan is currently in place for Irish Sea sole.

### 31.8 Uncertainties and bias in assessment and forecast

## Sampling

The deteriorating quality of the historic catch numbers-at-age data was considered to be a consequence of the low biological sampling intensity, and in particular the limited sampling in the first quarter. Therefore the combined age distribution was introduced in 2000 as an alternative method for raising the international catch numbers-at-age. The mean catch weights from this combined key were taken and the stock weights-at-age were obtained using a cohort interpolation method from the catch weights-at-age. Under the DCF there is an initiative to co-ordinate sampling across the three countries involved in the fishery. However, as the TAC is substantially reduced in the most recent years, sampling levels are also significantly reduced.

## Landings

There is no reliable information on the accuracy of the landing statistics. For the period 20052012, the total TAC uptake was only in the range of $50-98 \%$. In this context, misreporting was
not considered to be a major problem. In the most recent years, the TAC was substantially reduced and was restrictive in 2013 and 2014. In 2015-2018, $84 \%-90 \%$ of the TAC has been taken.

## Discards

The absence of discard data is unlikely to affect the quality of the assessment as information from recent years indicates that the average discarding by weight is $3.5 \%$ of the catch.

## Effort

There are no indications of Irish Sea sole fisheries misreporting effort. Effort in beam trawl fisheries that target sole has declined substantially in the last few years in accordance with the significant reductions in TAC.

## Surveys

The UK(E\&W)-BTS-Q3 survey appears to track year-class strength well. As previously investigated, this tuning fleet is also consistent in estimating year-class strength of the same year class at different ages. Therefore the Working Group had confidence in using the UK(E\&W)-BTS-Q3 survey as the only tuning fleet. The bias problem in the assessment maybe the result of the precise survey and less precise catch-at-age data.

## Model formulation

At present XSA is used to assess Irish Sea sole. In the WG of 2007 the model settings were changed which had a considerable impact on the estimates of SSB and fishing mortality. Due to these major revisions, ACFM changed the biomass reference points at its meeting of 2007. In the next two update assessments (2008-2009) no major changes were apparent. In the assessment of 2011, the settings were changed according to the outcome of WKFLAT 2011. The following assessments were update assessments. In 2016, the reference points were updated (see Section 33.5-33.6).

### 31.9 Recommendations for next Benchmark

The assessment diagnostics indicate a good correlation between the catch data and the survey tuning series. Therefore, at present there are no recommendations for a single stock Benchmark. However, in the recent years there has been great uncertainty from the fishing industry on the actual status of the sole stock in the Irish Sea. Fishermen are concerned that due to ecosystem changes and the changing fishing behaviour in the Irish Sea, science is no longer capturing the current situation. Because of this mismatch an EU action plan for the Irish Sea fisheries was set up. First, a comparative fishing study was suggested to compare the catch efficiency between the UK-BTS-Q3 and a Belgian commercial vessel. Secondly, a pilot industry-science beam trawl survey should reveal the spatial distribution of sole. The outcome of those work packages will indicate whether the data gathered by the UK-BTS-Q3 is still representative for the current situation or whether the implementation of an additional (annual) industry-science industry survey is needed. Thirdly, stock identification techniques (i.e. genetic fingerprinting and otolith shape analysis) will be performed to give insight on the origin and potential migration routes of sole that is caught in the Irish Sea.

The industry survey was not able to identify other areas of importance for sole in the Irish Sea than is already covered by the UK-BTS-Q3. Also, catchability and composition of catches in both surveys were comparable. These results suggest that the UK-BTS-Q3 gives a good representation of sole abundance and that an annual industry survey additional to this survey would not be of added value to the assessment. With regards to the stock identification study, the combination
of otolith shape analysis and genetic markers (SNPs) show subtle differences between the Irish Sea, Celtic Sea and Bristol Channel populations. However more samples from the different areas and from different years need to be analyzed to reveal what is driving these differences. Also, in the attempt to effectively reassign adult sole to their place of origin, it would be preferable to include a third stock identification technique: micro-chemical fingerprinting. Despite many questions yet unsolved, the pilot industry survey delivered valuable information that can be added to an ecosystem model for the Irish Sea (one of the aims of WKIrish: an ecosystem benchmark for the Irish Sea). Moreover, the survey was an example of a fruitful cooperation between fishermen and fisheries scientists and gave useful insights on how to cooperate with the fishing industry and to gain their trust in the collection of fisheries-independent data.

### 31.10 Management considerations

There is a stock-recruitment relationship for this stock and evidence of reduced recruitment at low levels of SSB. However, the recruitment for higher levels of SSB is less well defined (Figure 33.13).

Recruitment at-age 2 has been well below average since 2001. In the last four years, recruitment is estimated to be higher than the record low levels in 2011-2014. SSB is below Blim since 2004. XSA indicates that fishing mortality has fallen over the last couple of years (as did effort for most fleets fishing for Irish Sea sole), and is now well below Fmsy.

It is difficult for the stock to reach $\mathrm{B}_{\mathrm{pa}}$ MSY $\mathrm{B}_{\text {trigger }}$ in one year. A management plan for effort reduction that can be phased in over a number of years and implemented in conjunction with technical conservation measures should be considered.

Sole is caught in a mixed fishery with other flatfish as well as gadoids. Information from observer trips indicates that discarding of sole is relatively low.

### 31.11 Ecosystem considerations

Sole and plaice are primarily targeted by beam trawl fisheries. Beam trawling, is known to have an impact on the benthic communities, although less so on soft substrates and in areas which have been historically exploited by this fishing method. Some beam trawlers are using benthic drop-out panels that release about $75 \%$ of benthic invertebrates from the catches. Full square mesh codends are being tested in order to reduce the capture of benthos further and improve the selection profile of gadoids (Connolly, P.L. et al., 2009).

A complete ecosystem overview can be found in the stock annex Section A. 3

### 31.12 References

Connolly, P.L., Kelly, E., Dransfeld, L., Slattery, N., Paramor, O.A.L., and Frid, C.L.J. 2009. MEFEPO North Western Waters Atlas. Marine Institute.
ICES. 2014a. Report of the Workshop to consider reference points for all stocks (WKMSYREF2), 8-10 January 2014, ICES Headquarters, Copenhagen, Denmark. ICES CM 2014/ACOM:47. 91 pp.
ICES. 2014b. Report of the Joint ICES-MYFISH Workshop to consider the basis for Fmsy ranges for all stocks (WKMSYREF3), 17-21 November 2014, Charlottenlund, Denmark. ICES CM 2014/ACOM:64. 147 pp.

ICES. 2016b. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 183 pp.

ICES. 2016c. EU request to ICES to provide FMSY ranges for selected stocks in ICES subareas 5 to 10 , ICES special request advice. 5 February 2016 Version 2; 13 May 2016.

Table 29.1. Sol.27.7a - Nominal landings (tonnes) as officially reported by ICES, and working group estimates of the landings. Last year's landings are preliminary.
${ }^{1} 1989$ onwards: N. Ireland included with England \& Wales

| Year |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & E \\ & \frac{E}{D} \\ & \stackrel{D}{D} \\ & \infty \end{aligned}$ | $\begin{aligned} & \text { 쁜 } \\ & \text { 듄 } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { ס} \\ & \underline{\widetilde{T}} \\ & \underline{\underline{I N}} \end{aligned}$ |  | $\begin{aligned} & \underset{3}{3} \\ & \underset{\underset{y}{4}}{\underset{J}{4}} \end{aligned}$ |  |  |  |  |  |  | $\underset{1}{\bullet}$ |
| 1973 | 793 | 12 | 27 | 281 | 258 | - | 46 | 11 | 1428 | 0 | 1428 |  |
| 1974 | 664 | 54 | 28 | 320 | 218 | - | 23 | - | 1307 | 0 | 1307 |  |
| 1975 | 805 | 59 | 24 | 234 | 281 | - | 24 | 15 | 1442 | -1 | 1441 |  |
| 1976 | 674 | 72 | 74 | 381 | 195 | - | 49 | 18 | 1463 | 0 | 1463 |  |
| 1977 | 566 | 39 | 84 | 227 | 160 | - | 49 | 21 | 1146 | 1 | 1147 |  |
| 1978 | 453 | 65 | 127 | 177 | 189 | - | 57 | 30 | 1098 | 8 | 1106 |  |
| 1979 | 779 | 48 | 134 | 247 | 290 | - | 47 | 42 | 1587 | 27 | 1614 |  |
| 1980 | 1002 | 41 | 229 | 169 | 367 | - | 44 | 68 | 1920 | 21 | 1941 |  |
| 1981 | 884 | 13 | 167 | 186 | 311 | - | 41 | 45 | 1647 | 20 | 1667 |  |
| 1982 | 669 | 9 | 161 | 138 | 277 | - | 31 | 44 | 1329 | 9 | 1338 |  |
| 1983 | 544 | 3 | 203 | 224 | 219 | - | 33 | 29 | 1255 | -86 | 1169 |  |
| 1984 | 425 | 10 | 187 | 113 | 230 | - | 38 | 17 | 1020 | 38 | 1058 |  |
| 1985 | 589 | 9 | 180 | 546 | 269 | - | 36 | 28 | 1657 | -511 | 1146 |  |
| 1986 | 930 | 17 | 235 | - | 637 | 1 | 50 | 46 | 1916 | 79 | 1995 |  |
| 1987 | 987 | 5 | 312 | - | 599 | 3 | 72 | 63 | 2041 | 767 | 2808 | 2100 |
| 1988 | 915 | 11 | 366 | - | 507 | 1 | 47 | 38 | 1885 | 114 | 1999 | 1750 |
| 1989 | 1010 | 5 | 155 | - | 613 | 2 | . | 38 | 1823 | 10 | 1833 | 1480 |
| 1990 | 786 | 2 | 170 | - | 569 | 10 | . | 39 | 1576 | 7 | 1583 | 1500 |
| 1991 | 371 | 3 | 198 | - | 581 | 44 | . | 26 | 1223 | -11 | 1212 | 1500 |
| 1992 | 531 | 11 | 164 | - | 477 | 14 | . | 37 | 1234 | 25 | 1259 | 1350 |
| 1993 | 495 | 8 | 98 | - | 338 | 4 | . | 28 | 971 | 52 | 1023 | 1000 |
| 1994 | 706 | 7 | 226 | - | 409 | 5 | . | 14 | 1367 | 7 | 1374 | 1500 |
| 1995 | 675 | 5 | 176 | - | 424 | 12 | . | 8 | 1300 | -34 | 1266 | 1300 |
| 1996 | 533 | 5 | 133 | 149 | 194 | 4 | . | 5 | 1023 | -21 | 1002 | 1000 |


| Year |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \underline{E} \\ & \frac{1}{00} \\ & \hline 0 \end{aligned}$ |  | $\begin{aligned} & \mathbf{D} \\ & \underline{\pi} \\ & \underline{\underline{N}} \end{aligned}$ |  | $\underset{3}{3}$ $\underset{y}{\mathbf{3}}$ |  | $\begin{aligned} & \frac{7}{\bar{O}} \\ & \frac{\bar{C}}{0} \\ & \underline{\underline{0}} \\ & \dot{Z} \\ & \underline{y} \end{aligned}$ |  |  |  | Total used by WG | $\underset{\natural}{\text { U }}$ |
| 1997 | 570 | 3 | 130 | 123 | 189 | 5 | . | 7 | 1027 | -24 | 1003 | 1000 |
| 1998 | 525 | 3 | 134 | 60 | 161 | 3 | . | 9 | 895 | 16 | 911 | 900 |
| 1999 | 469 | <1 | 120 | 46 | 165 | 1 | . | 8 | 810 | 53 | 863 | 900 |
| 2000 | 493 | 3 | 135 | 60 | 133 | 1 | . | 8 | 833 | -15 | 818 | 1080 |
| 2001 | 674 | 4 | 135 | - | 195 | + | . | 4 | 1012 | 41 | 1053 | 1100 |
| 2002 | 817 | 4 | 96 | - | 165 | + | . | 3 | 1085 | 5 | 1090 | 1100 |
| 2003 | 687 | 4 | 103 | - | 217 | + | . | 3 | 1014 | 0 | 1014 | 1010 |
| 2004 | 527 | 1 | 77 | - | 106 | + | . | 1 | 712 | -3 | 709 | 800 |
| 2005 | 662 | 3 | 85 | - | 103 | + | . | 1 | 854 | 1 | 855 | 960 |
| 2006 | 419 | 1 | 85 | - | 69 | + | . | 2 | 576 | -7 | 569 | 960 |
| 2007 | 305 | 1 | 115 | - | 66 | <1 | . | 4 | 491 | 1 | 492 | 820 |
| 2008 | 216 | 1 | 66 | - | 37 | $\mathrm{n} / \mathrm{a}$ | . | n/a | 320 | 12 | 332 | 669 |
| 2009 | 257 | n/a | 47 | - | 19 | 1 | . | 1 | 325 | 0 | 325 | 502 |
| 2010 | 217 | <1 | 47 | - | 12 | <1 | . | n/a | 277 | 0 | 277 | 402 |
| 2011 | 250 | <1 | 48 | - | 31 | <1 | . | n/a | 330 | 0 | 330 | 390 |
| 2012 | 222 | <1 | 51 | - | 23 | <1 | - | n/a | 296 | 0 | 298 | 300 |
| 2013 | 96 | <1 | 40 | - | 12 | <1 | - | n/a | 148 | 0 | 148 | 140 |
| 2014 | 43 | $\mathrm{n} / \mathrm{a}$ | 43 | - | 10 | <1 | - | n/a | 96 | 0 | 99 | 95 |
| 2015 | 37 | n/a | 32 | - | 7 | n/a | - | n/a | 76 | 0 | 76 | 90 |
| 2016 | 14 | $\mathrm{n} / \mathrm{a}$ | 15 | - | 6 | n/a | - | n/a | 35 | 0 | 35 | 40 |
| 2017 | 14 | $\mathrm{n} / \mathrm{a}$ | 14 | - | 4 | n/a | - | n/a | 32 | 2 | 34 | 40 |
| 2018 | 14 | $\mathrm{n} / \mathrm{a}$ | 16 |  | 6 | n/a | - | n/a | 36 | 0 | 36 | 40 |

Table 29.2. Sol.27.7a - Catch numbers-at-age (in thousands).

| Age/Year | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 29 | 113 | 31 | 368 | 25 | 262 | 29 | 221 | 65 |
| 3 | 895 | 434 | 673 | 363 | 891 | 733 | 375 | 416 | 958 |
| 4 | 1009 | 2096 | 730 | 2195 | 576 | 2386 | 1331 | 1292 | 649 |
| 5 | 467 | 1130 | 1538 | 557 | 1713 | 539 | 2329 | 774 | 1009 |
| 6 | 1457 | 232 | 537 | 815 | 383 | 842 | 247 | 1066 | 442 |
| 7 | 289 | 878 | 172 | 267 | 422 | 157 | 544 | 150 | 638 |
| +gp | 2537 | 1886 | 1501 | 1143 | 971 | 1006 | 739 | 648 | 587 |
| Age/Year | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| 2 | 108 | 187 | 70 | 8 | 37 | 651 | 154 | 141 | 189 |
| 3 | 1027 | 940 | 580 | 346 | 165 | 786 | 1600 | 3334 | 3347 |
| 4 | 3432 | 1969 | 1668 | 1241 | 998 | 380 | 1085 | 3465 | 4104 |
| 5 | 829 | 3057 | 1480 | 1298 | 758 | 610 | 343 | 960 | 3184 |
| 6 | 637 | 521 | 1640 | 711 | 757 | 343 | 334 | 235 | 844 |
| 7 | 326 | 512 | 114 | 641 | 416 | 424 | 164 | 277 | 307 |
| +gp | 620 | 1146 | 865 | 397 | 709 | 557 | 739 | 848 | 808 |
| Age/Year | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 2 | 32 | 179 | 564 | 1316 | 363 | 83 | 122 | 132 | 60 |
| 3 | 444 | 771 | 1185 | 1269 | 2431 | 543 | 1343 | 920 | 469 |
| 4 | 4747 | 775 | 986 | 841 | 917 | 1965 | 1070 | 1444 | 1188 |
| 5 | 2100 | 3979 | 598 | 300 | 556 | 559 | 1579 | 737 | 741 |
| 6 | 1309 | 1178 | 2320 | 226 | 190 | 251 | 394 | 1010 | 430 |
| 7 | 203 | 552 | 592 | 1172 | 156 | 199 | 133 | 179 | 509 |
| +gp | 515 | 255 | 466 | 459 | 928 | 686 | 524 | 350 | 347 |
| Age/Year | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| 2 | 790 | 167 | 301 | 178 | 240 | 148 | 437 | 299 | 536 |
| 3 | 714 | 1728 | 1069 | 906 | 1438 | 930 | 825 | 862 | 1052 |
| 4 | 475 | 466 | 1259 | 907 | 822 | 1623 | 966 | 342 | 626 |


| Age/Year | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 711 | 256 | 297 | 600 | 717 | 740 | 795 | 368 | 271 |
| 6 | 409 | 315 | 115 | 150 | 511 | 575 | 302 | 304 | 314 |
| 7 | 258 | 191 | 136 | 55 | 80 | 254 | 217 | 139 | 279 |
| +gp | 532 | 423 | 232 | 258 | 272 | 217 | 345 | 181 | 368 |
| Age/Year | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| 2 | 112 | 171 | 99 | 92 | 22 | 17 | 17 | 23 | 12 |
| 3 | 670 | 356 | 353 | 414 | 336 | 225 | 148 | 99 | 49 |
| 4 | 649 | 348 | 190 | 333 | 233 | 401 | 311 | 75 | 59 |
| 5 | 203 | 243 | 195 | 146 | 177 | 176 | 274 | 106 | 37 |
| 6 | 113 | 86 | 156 | 132 | 65 | 97 | 116 | 78 | 38 |
| 7 | 151 | 41 | 56 | 127 | 72 | 54 | 52 | 34 | 51 |
| +gp | 379 | 298 | 209 | 162 | 158 | 122 | 115 | 82 | 56 |
| Age/Year | 2015 | 2016 | 2017 | 2018 |  |  |  |  |  |
| 2 | 15 | 1 | 2 | 4 |  |  |  |  |  |
| 3 | 36 | 18 | 41 | 22 |  |  |  |  |  |
| 4 | 37 | 22 | 19 | 46 |  |  |  |  |  |
| 5 | 30 | 14 | 15 | 14 |  |  |  |  |  |
| 6 | 17 | 10 | 5 | 9 |  |  |  |  |  |
| 7 | 21 | 7 | 6 | 3 |  |  |  |  |  |
| +gp | 74 | 32 | 13 | 10 |  |  |  |  |  |

Table 29.3. Sol.27.7a - Annual length distributions by country (2018).

|  | UK (England \& Wales) | Belgium | Ireland |
| :---: | :---: | :---: | :---: |
| Length (cm) | All gears | All gears | All gears |
| 21 |  |  |  |
| 22 |  |  |  |
| 23 | 54 |  |  |
| 24 | 215 | 908 | 185 |
| 25 | 430 | 3650 | 436 |
| 26 | 859 | 4971 | 586 |
| 27 | 1386 | 5850 | 1740 |
| 28 | 2106 | 6111 | 2068 |
| 29 | 1833 | 4685 | 2078 |
| 30 | 1118 | 5097 | 4756 |
| 31 | 613 | 3400 | 4272 |
| 32 | 871 | 2704 | 6051 |
| 33 | 796 | 2220 | 3143 |
| 34 | 344 | 1797 | 3872 |
| 35 | 161 | 1532 | 3925 |
| 36 | 420 | 1089 | 3382 |
| 37 | 134 | 865 | 2018 |
| 38 | 27 | 830 | 1437 |
| 39 | 0 | 566 | 828 |
| 40 | 183 | 787 | 887 |
| 41 |  | 225 | 306 |
| 42 |  | 246 | 335 |
| 43 |  | 80 | 316 |
| 44 |  | 130 | 83 |
| 45 |  | 154 | 33 |
| 46 |  | 51 | 33 |
| 47 |  | 0 |  |
| 48 |  | 22 |  |
| 49 |  | 15 |  |
| 50 |  |  |  |
| Total | 11550 | 47985 | 42770 |

Table 29.4. Sol.27.7a - Catch weights-at-age (kg).

| Age/Year | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.13 | 0.152 | 0.126 | 0.151 | 0.138 | 0.13 | 0.12 | 0.085 | 0.093 |
| 3 | 0.153 | 0.178 | 0.164 | 0.178 | 0.174 | 0.172 | 0.161 | 0.146 | 0.147 |
| 4 | 0.178 | 0.204 | 0.201 | 0.204 | 0.209 | 0.21 | 0.2 | 0.202 | 0.197 |
| 5 | 0.204 | 0.23 | 0.237 | 0.23 | 0.241 | 0.244 | 0.239 | 0.251 | 0.243 |
| 6 | 0.232 | 0.257 | 0.272 | 0.256 | 0.272 | 0.275 | 0.276 | 0.293 | 0.286 |
| 7 | 0.26 | 0.284 | 0.306 | 0.283 | 0.301 | 0.303 | 0.313 | 0.33 | 0.326 |
| +gp | 0.377 | 0.419 | 0.417 | 0.392 | 0.396 | 0.367 | 0.457 | 0.387 | 0.429 |
| Age/Year | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| 2 | 0.134 | 0.146 | 0.162 | 0.112 | 0.189 | 0.191 | 0.144 | 0.122 | 0.135 |
| 3 | 0.165 | 0.169 | 0.183 | 0.171 | 0.212 | 0.225 | 0.189 | 0.164 | 0.164 |
| 4 | 0.199 | 0.193 | 0.207 | 0.225 | 0.238 | 0.257 | 0.231 | 0.203 | 0.196 |
| 5 | 0.234 | 0.219 | 0.234 | 0.275 | 0.266 | 0.288 | 0.272 | 0.241 | 0.231 |
| 6 | 0.271 | 0.247 | 0.264 | 0.321 | 0.298 | 0.318 | 0.31 | 0.277 | 0.268 |
| 7 | 0.311 | 0.275 | 0.296 | 0.362 | 0.332 | 0.347 | 0.346 | 0.311 | 0.308 |
| +gp | 0.451 | 0.380 | 0.452 | 0.456 | 0.458 | 0.408 | 0.430 | 0.407 | 0.462 |
| Age/Year | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 2 | 0.111 | 0.125 | 0.135 | 0.133 | 0.149 | 0.102 | 0.175 | 0.129 | 0.156 |
| 3 | 0.147 | 0.163 | 0.162 | 0.172 | 0.177 | 0.156 | 0.198 | 0.182 | 0.193 |
| 4 | 0.183 | 0.201 | 0.192 | 0.208 | 0.207 | 0.205 | 0.227 | 0.232 | 0.228 |
| 5 | 0.218 | 0.237 | 0.227 | 0.241 | 0.239 | 0.248 | 0.261 | 0.277 | 0.263 |
| 6 | 0.252 | 0.271 | 0.265 | 0.272 | 0.274 | 0.285 | 0.301 | 0.318 | 0.296 |
| 7 | 0.286 | 0.304 | 0.307 | 0.3 | 0.31 | 0.318 | 0.346 | 0.356 | 0.327 |
| +gp | 0.419 | 0.389 | 0.414 | 0.345 | 0.379 | 0.370 | 0.509 | 0.451 | 0.410 |
| Age/Year | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| 2 | 0.154 | 0.187 | 0.179 | 0.14 | 0.175 | 0.162 | 0.16 | 0.17 | 0.16 |
| 3 | 0.197 | 0.209 | 0.217 | 0.189 | 0.18 | 0.172 | 0.187 | 0.219 | 0.203 |
| 4 | 0.237 | 0.234 | 0.252 | 0.25 | 0.271 | 0.211 | 0.247 | 0.289 | 0.256 |


| Age/Year | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 0.275 | 0.263 | 0.285 | 0.311 | 0.293 | 0.283 | 0.294 | 0.338 | 0.286 |
| 6 | 0.311 | 0.295 | 0.314 | 0.368 | 0.326 | 0.328 | 0.342 | 0.371 | 0.312 |
| 7 | 0.345 | 0.331 | 0.341 | 0.428 | 0.42 | 0.333 | 0.326 | 0.383 | 0.326 |
| +gp | 0.407 | 0.440 | 0.399 | 0.504 | 0.438 | 0.375 | 0.415 | 0.444 | 0.352 |
| Age/Year | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| 2 | 0.179 | 0.172 | 0.148 | 0.141 | 0.166 | 0.215 | 0.187 | 0.17 | 0.17 |
| 3 | 0.194 | 0.224 | 0.189 | 0.195 | 0.193 | 0.213 | 0.22 | 0.213 | 0.196 |
| 4 | 0.224 | 0.296 | 0.248 | 0.229 | 0.266 | 0.276 | 0.26 | 0.278 | 0.269 |
| 5 | 0.297 | 0.36 | 0.279 | 0.279 | 0.285 | 0.362 | 0.311 | 0.32 | 0.328 |
| 6 | 0.293 | 0.38 | 0.291 | 0.277 | 0.321 | 0.413 | 0.331 | 0.347 | 0.369 |
| 7 | 0.318 | 0.429 | 0.386 | 0.261 | 0.308 | 0.368 | 0.368 | 0.353 | 0.397 |
| +gp | 0.349 | 0.479 | 0.392 | 0.277 | 0.335 | 0.364 | 0.335 | 0.354 | 0.441 |
| Age/Year | 2015 | 2016 | 2017 | 2018 |  |  |  |  |  |
| 2 | 0.18 | 0.187 | 0.177 | 0.186 |  |  |  |  |  |
| 3 | 0.221 | 0.223 | 0.239 | 0.24 |  |  |  |  |  |
| 4 | 0.309 | 0.269 | 0.323 | 0.31 |  |  |  |  |  |
| 5 | 0.342 | 0.356 | 0.386 | 0.389 |  |  |  |  |  |
| 6 | 0.381 | 0.332 | 0.495 | 0.476 |  |  |  |  |  |
| 7 | 0.4 | 0.414 | 0.493 | 0.485 |  |  |  |  |  |
| +gp | 0.384 | 0.436 | 0.457 | 0.472 |  |  |  |  |  |

Table 29.5. Sol.27.7a - Stock weights-at-age (kg).

| Age/Year | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.13 | 0.152 | 0.126 | 0.151 | 0.138 | 0.13 | 0.12 | 0.085 | 0.093 |
| 3 | 0.153 | 0.178 | 0.164 | 0.178 | 0.174 | 0.172 | 0.161 | 0.146 | 0.147 |
| 4 | 0.178 | 0.204 | 0.201 | 0.204 | 0.209 | 0.21 | 0.2 | 0.202 | 0.197 |
| 5 | 0.204 | 0.23 | 0.237 | 0.23 | 0.241 | 0.244 | 0.239 | 0.251 | 0.243 |
| 6 | 0.232 | 0.257 | 0.272 | 0.256 | 0.272 | 0.275 | 0.276 | 0.293 | 0.286 |
| 7 | 0.26 | 0.284 | 0.306 | 0.283 | 0.301 | 0.303 | 0.313 | 0.33 | 0.326 |
| +gp | 0.377 | 0.419 | 0.417 | 0.392 | 0.396 | 0.367 | 0.457 | 0.387 | 0.429 |
| Age/Year | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| 2 | 0.134 | 0.146 | 0.162 | 0.112 | 0.189 | 0.191 | 0.144 | 0.122 | 0.135 |
| 3 | 0.165 | 0.169 | 0.183 | 0.171 | 0.212 | 0.225 | 0.189 | 0.164 | 0.164 |
| 4 | 0.199 | 0.193 | 0.207 | 0.225 | 0.238 | 0.257 | 0.231 | 0.203 | 0.196 |
| 5 | 0.234 | 0.219 | 0.234 | 0.275 | 0.266 | 0.288 | 0.272 | 0.241 | 0.231 |
| 6 | 0.271 | 0.247 | 0.264 | 0.321 | 0.298 | 0.318 | 0.31 | 0.277 | 0.268 |
| 7 | 0.311 | 0.275 | 0.296 | 0.362 | 0.332 | 0.347 | 0.346 | 0.311 | 0.308 |
| +gp | 0.451 | 0.380 | 0.452 | 0.456 | 0.458 | 0.408 | 0.430 | 0.407 | 0.462 |
| Age/Year | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 2 | 0.111 | 0.125 | 0.135 | 0.133 | 0.149 | 0.102 | 0.175 | 0.129 | 0.156 |
| 3 | 0.147 | 0.163 | 0.162 | 0.172 | 0.177 | 0.156 | 0.198 | 0.182 | 0.193 |
| 4 | 0.183 | 0.201 | 0.192 | 0.208 | 0.207 | 0.205 | 0.227 | 0.232 | 0.228 |
| 5 | 0.218 | 0.237 | 0.227 | 0.241 | 0.239 | 0.248 | 0.261 | 0.277 | 0.263 |
| 6 | 0.252 | 0.271 | 0.265 | 0.272 | 0.274 | 0.285 | 0.301 | 0.318 | 0.296 |
| 7 | 0.286 | 0.304 | 0.307 | 0.3 | 0.31 | 0.318 | 0.346 | 0.356 | 0.327 |
| +gp | 0.419 | 0.389 | 0.414 | 0.345 | 0.379 | 0.370 | 0.509 | 0.451 | 0.410 |
| Age/Year | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| 2 | 0.154 | 0.187 | 0.179 | 0.124 | 0.151 | 0.145 | 0.144 | 0.15 | 0.144 |
| 3 | 0.197 | 0.209 | 0.217 | 0.158 | 0.159 | 0.174 | 0.174 | 0.187 | 0.186 |
| 4 | 0.237 | 0.234 | 0.252 | 0.23 | 0.226 | 0.195 | 0.207 | 0.232 | 0.237 |


| Age/Year | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 0.275 | 0.263 | 0.285 | 0.303 | 0.271 | 0.277 | 0.249 | 0.289 | 0.288 |
| 6 | 0.311 | 0.295 | 0.314 | 0.345 | 0.318 | 0.31 | 0.311 | 0.331 | 0.325 |
| 7 | 0.345 | 0.331 | 0.341 | 0.41 | 0.393 | 0.33 | 0.327 | 0.362 | 0.348 |
| +gp | 0.407 | 0.440 | 0.399 | 0.530 | 0.450 | 0.397 | 0.383 | 0.419 | 0.383 |
| Age/Year | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| 2 | 0.152 | 0.156 | 0.134 | 0.129 | 0.158 | 0.167 | 0.156 | 0.149 | 0.111 |
| 3 | 0.177 | 0.2 | 0.181 | 0.17 | 0.165 | 0.188 | 0.218 | 0.2 | 0.183 |
| 4 | 0.213 | 0.24 | 0.236 | 0.208 | 0.228 | 0.231 | 0.235 | 0.248 | 0.24 |
| 5 | 0.276 | 0.284 | 0.288 | 0.263 | 0.256 | 0.31 | 0.293 | 0.288 | 0.302 |
| 6 | 0.289 | 0.336 | 0.324 | 0.278 | 0.3 | 0.343 | 0.346 | 0.329 | 0.343 |
| 7 | 0.315 | 0.354 | 0.383 | 0.276 | 0.292 | 0.344 | 0.39 | 0.342 | 0.371 |
| +gp | 0.348 | 0.419 | 0.424 | 0.319 | 0.305 | 0.340 | 0.345 | 0.358 | 0.399 |
| Age/Year | 2015 | 2016 | 2017 | 2018 |  |  |  |  |  |
| 2 | 0.153 | 0.127 | 0.152 | 0.149 |  |  |  |  |  |
| 3 | 0.194 | 0.2 | 0.212 | 0.206 |  |  |  |  |  |
| 4 | 0.246 | 0.244 | 0.268 | 0.272 |  |  |  |  |  |
| 5 | 0.303 | 0.332 | 0.322 | 0.355 |  |  |  |  |  |
| 6 | 0.353 | 0.337 | 0.42 | 0.429 |  |  |  |  |  |
| 7 | 0.384 | 0.397 | 0.405 | 0.49 |  |  |  |  |  |
| +gp | 0.397 | 0.411 | 0.443 | 0.462 |  |  |  |  |  |

Table 29.6a. Sol.27.7a - Discard rates for the main fleets operational in the Irish Sea (Belgian, UK and Irish beam trawl, UK and Irish otter trawl, UK and Irish Nephrops trawl).

|  | BEL | UK |  |  |  |  |  | IRL |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gear | TBB | TBB | OTB | TWIN OTB | NEPH OTB | TWIN NEPH | Other | TBB | NEPH OTB | OTB DEF |
| Landings ( t ) | 716 | 284 | 61 | 4 | 25 | 6 | Na | 427 | / | / |
| Discard ratio | 0.05 | 0.08 | 0.05 | 0.01 | 0.08 | 0.02 | Na | 0.02 | / | / |
| years | 2007-2009 | $\begin{gathered} 2002 \\ 2005-2007 \end{gathered}$ | 2002-2009 | 2003,2004,2007 | $\begin{gathered} 2003, \\ 2006-2009 \end{gathered}$ | 2002,2003,2008 | Na | 2003-2009 | / | / |
| Landings (t) 2010 | 210.917 | 1.721 | 1.071 | 0.014 | 3.329 | 0.501 | 0.741 | 38.283 | 5.327 | 3.632 |
| Discard ratio 2010 | 0.04 | Na | 0.00 | Na | 0.05 | Na | Na | 0.05 | 0.16* | 0.39* |
| Landings (t) 2011 | 239.483 | 13.662 | 2.866 | 0.05 | 5.201 | 0.414 | 0.821 | 32.514 | 10.116 | 5.581 |
| Discard ratio 2011 | 0.04 | Na | 0.02 | Na | 0.00 | Na | Na | 0.003 | 0.16* | 0.00 |

* It should be noted that the $\mathbf{1 6 \%}$ discard rate for 2010-2011 of the Irish Nephrops fleet and the $\mathbf{3 9 \%}$ discard rate for 2010 of the Irish otter trawl fleet only accounts for respectively $1.9 \%$, $3.1 \%$ and $1.3 \%$ of the total international landings.

Table 29.6b. Sol.27.7a - Discard rates.

| Country | Year | Landings (L) (t) |  |  | Discards (D) (t) |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | TBB | ОтB | other |  |
| BE | 2012 | 213.392 | 8.301 | 0 | 16.222 |
|  | 2013 | 93.009 | 3.028 | 0 | 8.538 |
|  | 2014 | 36.144 | 7.288 | 0 | 2.286 |
|  | 2015 | 32.2 | 3.995 | 0 | 2.343 |
|  | 2016 | 12.533 | 1.538 | 0 | 0.336 |
|  | 2017 | 11.047 | 2.154 | 0 | 0.436 |
|  | 2018 | 13.2 | 1.085 | 0 | 0.5 |
| UK | 2012 | 7.278 | 5.459 | 1.229 | 0 |
|  | 2013 | 0.168 | 5.108 | 1.258 | 0 |
|  | 2014 | 0.149 | 3.579 | 1.582 | 1.404 |
|  | 2015 | 0.164 | 3.505 | 0.491 | 0 |
|  | 2016 | 0.110 | 2.700 | 0.641 | 0.029 |
|  | 2017 | 0.06 | 1.449 | 1.004 | 0 |
|  | 2018 | 0.099 | 2.259 | 0.877 | 0.000 |
| IR | 2012 | 38.79 | 8.162 | 3.824 | 1 |
|  | 2013 | 30.934 | 9.23 | 0.009 | 0 |



Table 29.7a. Sol.27.7a - Effort series.

| Year | Belgium <br> beam ${ }^{1}$ <br> Whole year | UK(E\&W) <br> beam ${ }^{2}$ <br> Whole year | beam ${ }^{3}$ <br> Whole year | otter ${ }^{2}$ <br> Whole year | otter ${ }^{3}$ <br> Whole year | Ireland <br> otter ${ }^{4}$ <br> Whole year | beam ${ }^{4}$ <br> Whole year |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
| 1972 | - | - | - | 128.4 | - | - | - |
| 1973 | - | - | - | 147.6 | - | - | - |
| 1974 | - | - | - | 115.2 | - | - | - |
| 1975 | 28.4 | - | - | 130.7 | - | - | - |
| 1976 | 24.9 | - | - | 122.3 | - | - | - |
| 1977 | 22.1 | - | - | 101.9 | - | - | - |
| 1978 | 17.5 | 0.9 | - | 89.1 | - | - | - |
| 1979 | 20.4 | 1.7 | - | 89.9 | - | - | - |
| 1980 | 32.0 | 4.3 | - | 107.0 | - | - | - |
| 1981 | 36.5 | 6.4 | - | 107.1 | - | - | - |
| 1982 | 26.5 | 5.5 | - | 127.2 | - | - | - |
| 1983 | 28.7 | 2.8 | 0.0 | 88.1 | 1716.5 | - | - |
| 1984 | 17.5 | 4.1 | 263.0 | 103.1 | 7932.1 | - | - |
| 1985 | 27.0 | 7.4 | 428.1 | 102.9 | 6930.8 | - | - |
| 1986 | 44.5 | 17.0 | 1122.9 | 90.3 | 6693.2 | - | - |
| 1987 | 51.6 | 22.0 | 1178.5 | 130.6 | 9008.9 | - | - |
| 1988 | 38.2 | 18.6 | 1019.2 | 132.0 | 8292.4 | - | - |
| 1989 | 42.2 | 25.3 | 1344.5 | 139.5 | 16161.4 | - | - |
| 1990 | 42.4 | 31.0 | 1473.1 | 117.1 | 7724.5 | - | - |
| 1991 | 17.1 | 25.8 | 1211.3 | 107.3 | 7081.1 | - | - |
| 1992 | 25.1 | 23.4 | 908.1 | 96.8 | 6671.8 | - | - |
| 1993 | 23.9 | 21.5 | 826.9 | 78.9 | 6013.1 | - | - |
| 1994 | 32.5 | 20.1 | 1451.6 | 43.0 | 3060.0 | - | - |
| 1995 | 28.6 | 20.9 | 1429.4 | 43.1 | 3357.0 | 80.3 | 8.64 |
| 1996 | 23.2 | 13.3 | 894.3 | 42.2 | 3085.1 | 64.8 | 6.26 |
| 1997 | 30.7 | 10.8 | 784.4 | 39.9 | 2903.3 | 92.2 | 9.86 |


| Year | Belgium <br> beam ${ }^{1}$ <br> Whole year | UK(E\&W) <br> beam ${ }^{2}$ <br> Whole year | beam ${ }^{3}$ <br> Whole year | otter ${ }^{2}$ <br> Whole year | otter ${ }^{3}$ <br> Whole year | Ireland <br> otter ${ }^{4}$ <br> Whole year | beam ${ }^{4}$ <br> Whole year |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
| 1998 | 24.7 | 10.4 | 696.0 | 36.9 | 2620.6 | 93.5 | 11.58 |
| 1999 | 22.7 | 11.0 | 778.9 | 22.9 | 1803.5 | 110.3 | 14.7 |
| 2000 | 26.0 | 6.3 | 410.7 | 27.0 | 2034.9 | 82.7 | 11.4 |
| 2001 | 36.8 | 12.5 | 767.4 | 32.8 | 2352.9 | 77.5 | 13.1 |
| 2002 | 47.0 | 8.0 | 535.1 | 24.8 | 1774.0 | 77.9 | 17.7 |
| 2003 | 43.6 | 14.0 | 863.7 | 23.9 | 1728.3 | 73.9 | 18.7 |
| 2004 | 32.0 | 7.4 | 419.9 | 23.5 | 1727.0 | 72.5 | 14.2 |
| 2005 | 37.5 | 11.4 | 627.8 | 16.7 | 1313.6 | 68.3 | 14.7 |
| 2006 | 24.6 | 4.6 | 280.1 | 5.2 | 478.5 | 66.2 | 12.2 |
| 2007 | 19.4 | 3.2 | 193.5 | 4.4 | 397.2 | 74.1 | 14.2 |
| 2008 | 9.6 | 1.3 | 98.0 | 2.7 | 320.4 | 58.8 | 9.5 |
| 2009 | 11.1 | 0.5 | 24.9 | 1.5 | 157.7 | 42.8 | 7.6 |
| 2010 | 11.1 | 0.2 | 10.2 | 1.4 | 151.0 | 45.8 | 9.4 |
| 2011 | 12.5 | 1.6 | 91.2 | 0.7 | 72.7 | 54.5 | 8.1 |
| 2012 | 10.9 | 0.9 | 60.7 | 0.4 | 85.0 | 58.3 | 7.2 |
| 2013 | 7.0 | 0.0 | 1.3 | 0.3 | 31.9 | 42.6 | 5.0 |
| 2014 | 3.9 | - | 0.4 | - | 16.1 | 47.7 | 6.0 |
| 2015 | 3.5 | - | 0.9 | - | 0.0 | 39.8 | 8.3 |
| 2016 | 1.8 | - | 3.9 | - | 0.0 | 33.4 | 7.9 |
| 2017 | 3.0 | - | 0.0 | - | 160.7 | 12.1 | 7.5 |
| 2018 | 2.5 | - | 0.0 | - | 238.1 | 13.6 | 9.6 |

All the trawlers fishing in the Irish Sea (UK fleet) are below 12 meters in length.
${ }^{1} 000$ hours fishing.
${ }^{2} 000$ 'hours fished (GRT corrected $>40$ ' vessels).
${ }^{3}$ days fished.
${ }^{4} 000$ 'hours.
${ }^{7}$ days fished.

* Provisional.

Table 29.7b. Sol.27.7a - Lpue.

| Year | Belgium <br> beam ${ }^{1}$ <br> Whole year | UK(E\&W) <br> beam ${ }^{3}$ <br> Whole year | beam ${ }^{2}$ <br> Whole year | otter ${ }^{3}$ <br> Whole year | otter ${ }^{2}$ <br> Whole year | UK <br> beam survey ${ }^{4}$ |  | Ireland <br> otter <br> Whole year | beam <br> Whole year |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  | Sept | Marc h |  |  |
| 1972 | - | - | - | 1.06 | - | - | - | - | - |
| 1973 | - | - | - | 1.06 | - | - | - | - | - |
| 1974 | - | - | - | 1.09 | - | - | - | - | - |
| 1975 | 21.4 | - | - | 1.39 | - | - | - | - | - |
| 1976 | 23.1 | - | - | 0.94 | - | - | - | - | - |
| 1977 | 19.8 | - | - | 0.80 | - | - | - | - | - |
| 1978 | 18.1 | 34.32 | - | 1.04 | - | - | - | - | - |
| 1979 | 33.4 | 32.01 | - | 1.43 | - | - | - | - | - |
| 1980 | 28.2 | 31.70 | - | 1.01 | - | - | - | - | - |
| 1981 | 22.2 | 21.32 | - | 0.75 | - | - | - | - | - |
| 1982 | 22.0 | 29.94 | - | 0.53 | - | - | - | - | - |
| 1983 | 13.9 | 37.31 | 0.0 | 0.57 | 150.2 | - | - | - | - |
| 1984 | 22.5 | 16.24 | 2851.4 | 0.71 | 119.3 | - | - | - | - |
| 1985 | 20.6 | 17.34 | 2956.3 | 0.56 | 135.7 | - | - | - | - |
| 1986 | 19.1 | 19.23 | 3925.7 | 0.84 | 174.9 | - | - | - | - |
| 1987 | 17.7 | 14.82 | 3726.9 | 0.77 | 144.9 | - | - | - | - |
| 1988 | 21.3 | 11.81 | 2673.3 | 0.46 | 80.3 | 161.9 | - | - | - |
| 1989 | 21.9 | 9.17 | 1750.6 | 0.70 | 138.9 | 150.0 | - | - | - |
| 1990 | 17.5 | 9.52 | 2300.9 | 0.61 | 119.7 | 196.9 | - | - | - |
| 1991 | 18.7 | 10.43 | 2420.9 | 1.12 | 177.4 | 175.7 | - | - | - |
| 1992 | 19.2 | 9.50 | 2763.0 | 1.02 | 126.0 | 162.6 | - | - | - |
| 1993 | 20.0 | 7.60 | 1879.8 | 0.54 | 69.1 | 100.1 | 104.7 | - | - |
| 1994 | 19.1 | 11.76 | 1479.9 | 0.74 | 88.1 | 110.7 | 91.9 |  | - |
| 1995 | 18.1 | 14.96 | 1721.1 | 0.95 | 142.3 | 92.04 | 79.3 | 0.38 | 12.69 |
| 1996 | 17.7 | 9.44 | 1471.7 | 0.53 | 47.7 | 89.48 | - | 0.25 | 14.94 |
| 1997 | 16.6 | 10.49 | 961.8 | 0.73 | 103.2 | 155.7 | 63.3 | 0.23 | 8.53 |
| 1998 | 19.0 | 8.42 | 907.8 | 0.48 | 50.5 | 144.9 | 89.3 | 0.38 | 7.77 |
| 1999 | 19.5 | 9.94 | 1124.9 | 0.60 | 64.8 | 116.0 | - | 0.29 | 9.22 |
| 2000 | 15.5 | 12.90 | 1604.7 | 0.44 | 34.6 | 130.7 | - | 0.29 | 8.49 |
| 2001 | 15.0 | 11.72 | 1537.4 | 0.15 | 23.4 | 96.87 | - | 0.38 | 7.86 |
| 2002 | 15.0 | 16.73 | 1484.3 | 1.48 | 98.8 | 76.73 | - | 0.32 | 4.67 |
| 2003 | 14.8 | 13.20 | 1351.6 | 0.15 | 340.4 | 88.55 | - | 0.34 | 4.20 |
| 2004 | 15.4 | 13.86 | 941.7 | 0.17 | 27.6 | 98.92 | - | 0.14 | 4.31 |
| 2005 | 16.7 | 9.14 | 1199.9 | 0.19 | 21.3 | 48.91 | - | 0.16 | 4.70 |
| 2006 | 15.2 | 7.83 | 826.1 | 0.52 | 34.8 | 52.63 | - | 0.16 | 6.00 |


| Year | Belgium | UK(E\&W) <br> beam ${ }^{3}$ |  |  |  | UK |  | Ireland |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | beam ${ }^{1}$ |  | beam ${ }^{2}$ | otter ${ }^{3}$ | otter ${ }^{2}$ | beam survey ${ }^{4}$ |  | otter | beam |
|  | Whole year | Whole year | Whole year | Whole year | Whole year | Sept | Marc h | Whole year | Whole year |
| 2007 | 13.7 | 16.38 | 1629.9 | 0.42 | 21.4 | 53.05 | - | 0.37 | 6.37 |
| 2008 | 19.5 | 15.25 | 887.4 | 0.30 | 16.4 | 50.67 | - | 0.20 | 6.08 |
| 2009 | 20.2 | 18.88 | 1201.2 | 0.22 | 13.6 | 45.75 | - | 0.28 | 4.53 |
| 2010 | 18.0 | 13.90 | 262.3 | 0.46 | 17.8 | 27.80 | - | 0.19 | 4.09 |
| 2011 | 17.6 | 4.45 | 322.5 | 0.18 | 13.7 | 36.97 | - | 0.30 | 4.13 |
| 2012 | 18.9 | 4.27 | 99.9 | 0.08 | 10.5 | 26.47 | - | 0.14 | 5.41 |
| 2013 | 12.7 | - | 27.7 | 0.10 | 3.4 | 31.65 | - | 0.22 | 6.27 |
| 2014 | 8.9 | - | 0.0 | - | 0.0 | 41.14 | - | 0.14 | 5.40 |
| 2015 | 8.9 | - | 146.1 | - | 0.0 | 58.88 | - | 0.18 | 3.14 |
| 2016 | 6.5 | - | 0.0 | - | 0.0 | 69.35 | - | 0.18 | 1.17 |
| 2017 | 3.6 | - | 0.0 | - | 5.6 | 64.24 | - | 0.36 | 1.23 |
| 2018 | 5.4 | - | 0.0 | - | 12.6 | 78.51 | - | 0.28 | 1.49 |

All lpue values in $\mathrm{Kg} / \mathrm{hr}$.
${ }^{1} \mathrm{Kg} / 000$ 'hr.
${ }^{2} \mathrm{Kg} /$ day.
${ }^{3} \mathrm{Kg} / 000$ 'hr fished (GRT corrected $>40$ ' vessels).
${ }^{4} \mathrm{Kg} / 100 \mathrm{~km}$ fished.

* Provisional.

Table 29.8. Sol.27.7a - Tuning series (values in bold are used in the assessment).


Table 29.8. Sole in 7.a - Continued (values in bold are used in the assessment).


Table 29.8. Sole in 7.a. Continued (values in bold are used in the assessment).

| UK(E\&W)-COT |  | UK Commercial Otter trawl |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 2013 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2 | 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| 107.3 | 265 | 155 | 63 | 29 | 19 | 71 | 20 | 11 | 2 | 0 | 1 | 1 | 1 |
| 96.8 | 16 | 224 | 69 | 22 | 16 | 10 | 36 | 10 | 10 | 1 | 0 | 0 | 0 |
| 78.9 | 9 | 27 | 77 | 19 | 3 | 7 | 4 | 5 | 1 | 2 | 0 | 0 | 0 |
| 43 | 4 | 66 | 34 | 50 | 20 | 3 | 4 | 4 | 7 | 1 | 2 | 0 | 0 |
| 43.1 | 17 | 50 | 34 | 15 | 24 | 7 | 1 | 2 | 0 | 2 | 1 | 1 | 0 |
| 42.2 | 2 | 5 | 18 | 12 | 7 | 12 | 4 | 1 | 1 | 1 | 1 | 1 | 1 |
| 39.9 | 14 | 15 | 7 | 14 | 9 | 3 | 7 | 3 | 1 | 1 | 0 | 1 | 0 |
| 36.9 | 5 | 24 | 5 | 3 | 5 | 3 | 2 | 2 | 1 | 1 | 0 | 0 | 0 |
| 22.8 | 5 | 15 | 12 | 2 | 0 | 2 | 1 | 1 | 1 | 1 | 0 | 0 | 0 |
| 27 | 2 | 12 | 9 | 8 | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 32.9 | 3 | 10 | 6 | 8 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 24.8 | 0 | 8 | 16 | 3 | 5 | 3 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
| 23.9 | 1 | 2 | 6 | 4 | 2 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| 23.5 | 3 | 5 | 3 | 4 | 3 | 2 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 16.7 | 2 | 4 | 2 | 1 | 2 | 2 | 1 | 1 | 1 | 0 | 0 | 0 | 0 |
| 5.2 | 1 | 2 | 4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 0 | 0 | 0 |
| 4.4 | 1 | 1 | 2 | 2 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 |
| 2.7 | 0 | 1 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1.54 | 0 | 0 | 0.2 | 0.3 | 0.1 | 0.2 | 0.2 | 0 | 0 | 0.1 | 0 | 0 | 0 |
| 1.42 | 0 | 0.1 | 0.2 | 0.3 | 0.1 | 0.1 | 0.2 | 0.1 | 0 | 0.1 | 0.1 | 0.1 | 0 |
| 0.686 | 0 | 0.1 | 0.1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.241 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.272 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| IR-COT | Irish Commercial Otter trawl |  |  |  |  |  |  |  |  |  |  |  |  |
| 1995 | 2005 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2 | 10 |  |  |  |  |  |  |  |  |  |  |  |  |
| 70682 | 6.8 | 17.7 | 25.5 | 9.2 | 25.8 | 3.6 | 0.8 | 1.5 | 1.9 | 1995 |  |  |  |
| 58166 | 0 | 5.7 | 12.9 | 12.7 | 4.7 | 4.7 | 2.2 | 0.2 | 0 | 1996 |  |  |  |
| 75029 | 27.8 | 10.2 | 4.1 | 9.2 | 6.4 | 3.5 | 3.9 | 1 | 0.2 | 1997 |  |  |  |
| 81073 | 5.5 | 40.7 | 14.7 | 6.6 | 12.3 | 5.4 | 2.7 | 4.1 | 1 | 1998 |  |  |  |
| 93221 | 26.6 | 36.8 | 30.9 | 5.1 | 3.8 | 5.3 | 2.4 | 0.5 | 1.2 | 1999 |  |  |  |
| 64320 | 1.6 | 13.2 | 13.4 | 11 | 3.4 | 1.1 | 1 | 0.4 | 0 | 2000 |  |  |  |
| 77541 | 0.2 | 6.1 | 18.6 | 18.6 | 10.8 | 2.1 | 4.1 | 1.3 | 0.3 | 2001 |  |  |  |
| 39996 | 20.3 | 20 | 30.2 | 16.4 | 8.2 | 2.9 | 2.4 | 1.4 | 0.5 | 2002 |  |  |  |
| 73854 | 0.9 | 35.9 | 21.7 | 9.8 | 3.3 | 0.5 | 0.8 | 0.2 | 0.2 | 2003 |  |  |  |
| 72507 | 9 | 15.1 | 4.1 | 3.2 | 1.9 | 1.6 | 0.3 | 0.2 | 0.1 | 2004 |  |  |  |
| \#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\# |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 31142 | 4 | 1.7 | 1.6 | 1.6 | 0.6 | 0.1 | 0 | 0 | 0 | 2005 |  |  |  |

Table 29.9. Sol.27.7a - Diagnostics.

```
FLR XSA Diagnostics 2019-05-21 10:57:56
CPUE data from indices
Catch data for 49 years. 1970 to 2018. Ages 2 to 8.
    fleet first age last age first year last year alpha beta
1 UK (E&W)-BTS-Q3 2 % 7 1988 2018 0.75 0.85
Time series weights :
    Tapered time weighting not applied
Catchability analysis :
    Catchability independent of size for all ages
    Catchability independent of age for ages > 4
Terminal population estimation :
    Survivor estimates shrunk towards the mean F
    of the final }5\mathrm{ years or the 3 oldest ages.
    S.E. of the mean to which the estimates are shrunk = 1.5
    Minimum standard error for population
    estimates derived from each fleet = 0.3
    prior weighting not applied
Regression weights
        year
age 2009 2010 2011 2012 2013 2014 2015 2016 2017 2018
    all 1 1 1 1 1 1 
    Fishing mortalities
    year
age 2009 2010}202011 2012 2013 2014 2015 2016 2017 2018
    2 0.043 0.014 0.028 0.020 0.038 0.015 0.009 0.000 0.001 0.001
    3 0.299 0.197 0.179 0.325 0.143 0.096 0.053 0.012 0.011 0.017
    4 0.405 0.245 0.339 0.357 0.242 0.107 0.088 0.036 0.014 0.014
    5 0.476 0.348 0.263 0.363 0.176 0.162 0.066 0.038 0.030 0.011
    6 0.263 0.357 0.290 0.248 0.148 0.080 0.094 0.025 0.016 0.020
    7 0.387 0.201 0.500 0.223 0.096 0.123 0.052 0.045 0.017 0.011
    8 0.387 0.201 0.500 0.223 0.096 0.123 0.052 0.045 0.017 0.011
    XSA population number (Thousand)
        age
year 
    2009 2278 1681 1050 405 599 416 528
    2010}161519741128 633 228 417 912
    2011
    2012
    2013
    2014}88315564 614 261 523 464 509
    2015}180
    20164422 1621
    2017 1553 4000 1450 555 335 374 809
    2018}36701403 35801293487 298 993
    Estimated population abundance at 1st Jan 2019
        age
    year }\begin{array}{llllllll}{2}&{3}&{4}&{5}&{6}&{7}&{8}
    20190 331812493198 1158433 267
```

Log catchability residuals.

| year |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |  |
| 2 | 0.044 | 0.027 | 0.411 | 0.509 | -0.051 | -0.273 | 0.162 | 0.178 | -0.279 | 0.095 |  |
| 3 | 0.587 | 0.369 | -0.124 | -0.292 | 0.473 | -0.271 | -0.043 | 0.295 | -0.674 | -0.072 |  |
| 4 | 0.007 | 0.070 | -0.240 | -0.923 | 0.451 | -0.098 | -0.284 | 0.051 | -0.245 | -0.161 |  |
| 5 | -0.383 | -0.018 | 0.969 | -0.611 | -0.017 | -0.312 | 0.031 | -0.579 | -0.219 | 0.033 |  |
| 6 | -0.231 | -0.234 | 0.302 | -0.198 | 0.173 | -0.068 | 0.540 | -0.017 | -0.172 | -0.156 |  |
| 7 | -0.118 | 0.086 | 0.192 | -0.190 | -0.195 | -0.077 | 0.192 | -0.337 | -0.149 | 0.281 |  |
| year |  |  |  |  |  |  |  |  |  |  |  |
| age | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |  |
| 2 | 0.443 | -0.150 | 0.000 | -0.043 | -0.901 | 0.139 | 0.040 | 0.000 | 0.276 | -0.234 |  |
| 3 | 0.110 | 0.007 | -0.207 | -0.222 | -0.225 | -0.173 | 0.421 | -0.365 | 0.149 | 0.251 |  |
| 4 | -0.765 | 0.322 | 0.325 | -0.484 | 0.070 | 0.237 | -0.098 | -0.216 | -0.091 | 0.266 |  |
| 5 | -0.753 | 0.339 | -0.116 | -0.137 | -0.387 | 0.206 | 0.447 | -0.065 | 0.720 | 0.268 |  |
| 6 | -0.279 | 0.360 | 0.155 | -0.094 | 0.077 | 0.005 | 0.043 | 0.177 | 0.250 | -0.028 |  |
| 7 | 0.206 | 0.189 | -0.116 | -0.014 | -0.013 | -0.230 | 0.355 | -0.024 | -0.203 | -0.025 |  |
| year |  |  |  |  |  |  |  |  |  |  |  |
| age | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| 2 | 0.015 | 0.002 | -0.593 | -0.163 | -0.158 | 0.287 | -0.262 | 0.206 | 0.218 | -0.044 | 0.097 |
| 3 | 0.028 | 0.048 | -0.110 | 0.044 | -0.251 | 0.051 | -0.152 | 0.164 | 0.175 | -0.093 | 0.100 |
| 4 | 0.050 | 0.210 | -0.068 | 0.383 | 0.304 | 0.487 | 0.374 | 0.080 | 0.125 | -0.089 | -0.049 |
| 5 | 0.395 | 0.475 | -0.186 | 0.266 | -0.159 | 0.255 | 0.360 | 0.128 | 0.258 | 0.037 | -0.365 |
| 6 | 0.128 | 0.370 | -0.379 | -0.103 | -0.454 | -0.226 | 0.215 | -0.462 | 0.333 | 0.061 | 0.025 |
| 7 | -0.202 | -0.082 | -0.521 | -0.001 | 0.147 | 0.302 | 0.156 | -0.177 | 0.192 | -0.558 | -0.338 |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

|  | 2 | 3 | 4 | 5 | 6 | 7 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean_Logq | -7.4440 | -7.7724 | -7.9097 | -7.9097 | -7.9097 | -7.9097 |
| S.E_Logq | 0.2943 | 0.2943 | 0.2943 | 0.2943 | 0.2943 | 0.2943 |

Terminal year survivor and $F$ summaries:
Age = 2 . Catchability constand w.r.t. time and dependant on age
Year class = 2016
Fleet $=$ fshk
$\begin{array}{lr} & 2 \\ \text { Survivors } & 294.000 \\ \text { Raw weights } & 0.444\end{array}$

Fleet $=$ UK (E\&W)-BTS-Q3
Survivors 3656.000
Raw weights 11.098


Weighted prediction:

| Suvivors Int.s.e. Ext.s.e. Var.Ratio F |  |
| :--- | :--- |
| $[1$,$] "3318" ""$ | " " |

Age $=3$. Catchability constand w.r.t. time and dependant on age
Year class = 2015
Fleet $=$ fshk


Weighted prediction:
Suvivors Int.s.e. Ext.s.e. Var.Ratio F
[1,] "1158" "" "" "" "0.011"

Age $=6$. Catchability constand w.r.t. time and dependant on age

```
Year class = 2012
Fleet = fshk
\begin{tabular}{lr} 
& 6 \\
Survivors & 113.000 \\
Raw weights & 0.444
\end{tabular}
\begin{tabular}{lrrrrr} 
Fleet \(=\) UK (E\&W)-BTS-Q3 \\
& 6 & 5 & 4 & 3 & 2 \\
Survivors & 443.000 & 449.000 & 490.000 & 510.000 & 333.00 \\
Raw weights & 10.896 & 6.063 & 8.283 & 9.678 & 9.53
\end{tabular}
```



Weighted prediction:
Suvivors Int.s.e. Ext.s.e. Var.Ratio F
[1,] "433" "" "" "" "0.02"

Age = 7 . Catchability constand w.r.t. time and dependant on age Year class = 2011

Fleet = fshk
Survivors 190.000
Raw weights 0.444

|  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Fleet $=$ UK $(E \& W)-$ BTS-Q3 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |
|  | 7 | 4 | 3 | 2 |  |  |
| Survivors | 190.000 | 284.000 | 345.000 | 289.00 | 229.00 | 356.000 |
| Raw weights | 10.994 | 10.817 | 5.966 | 7.74 | 8.66 | 8.337 |



Weighted prediction:
Suvivors Int.s.e. Ext.s.e. Var.Ratio F
[1,] "267" "" "" "" "0.011"

Table 29.10. Sol.27.7a - Fishing mortality.

| Age/Year |  | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 |  | 0.0083 | 0.0117 | 0.0103 | 0.0299 | 0.0045 | 0.0421 | 0.0079 | 0.0148 | 0.0076 | 0.0129 | 0.0396 | 0.0165 |
| 3 |  | 0.1196 | 0.1480 | 0.0810 | 0.1436 | 0.0847 | 0.1575 | 0.0704 | 0.1350 | 0.0743 | 0.1427 | 0.1335 | 0.1488 |
| 4 |  | 0.2956 | 0.3987 | 0.3520 | 0.3621 | 0.3158 | 0.3033 | 0.4192 | 0.3255 | 0.2866 | 0.3645 | 0.3931 | 0.3289 |
| 5 |  | 0.4445 | 0.5543 | 0.5060 | 0.4394 | 0.4723 | 0.4845 | 0.4816 | 0.4072 | 0.4036 | 0.6321 | 0.5673 | 0.5109 |
| 6 |  | 0.4292 | 0.3670 | 0.4932 | 0.4874 | 0.5436 | 0.3973 | 0.3792 | 0.3752 | 0.3815 | 0.4260 | 0.9493 | 0.6028 |
| 7 |  | 0.3909 | 0.4415 | 0.4519 | 0.4310 | 0.4454 | 0.3962 | 0.4281 | 0.3704 | 0.3583 | 0.4759 | 0.6393 | 0.4826 |
| +gp |  | 0.3909 | 0.4415 | 0.4519 | 0.4310 | 0.4454 | 0.3962 | 0.4281 , | 0.3704 | 0.3583 | 0.4759 | 0.6393 | 0.4826 |
| FBAR 4-7 |  | 0.3901 | 0.4404 | 0.4508 | 0.4300 | 0.4443 | 0.3953 | 0.4270 | 0.3696 | 0.3575 | 0.4746 | 0.6373 | 0.4813 |
| Age/Year |  | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 2 |  | 0.0034 | 0.0070 | 0.0452 | 0.0100 | 0.0063 | 0.0591 | 0.0097 | 0.0440 | 0.1127 | 0.1155 | 0.0801 | 0.0142 |
| 3 |  | 0.0953 | 0.0813 | 0.1808 | 0.1343 | 0.2755 | 0.1800 | 0.1719 | 0.2992 | 0.3990 | 0.3520 | 0.2876 | 0.1483 |
| 4 |  | 0.4772 | 0.3835 | 0.2429 | 0.3602 | 0.4215 | 0.5647 | 0.3699 | 0.4498 | 0.6796 | 0.4849 | 0.4109 | 0.3532 |
| 5 |  | 0.4078 | 0.5321 | 0.3794 | 0.3205 | 0.5519 | 0.7603 | 0.5606 | 0.5354 | 0.6625 | 0.3958 | 0.6085 | 0.4186 |
| 6 |  | 0.4370 | 0.3922 | 0.4328 | 0.3273 | 0.3370 | 1.2558 | 0.7284 | 0.6275 | 0.6094 | 0.4978 | 0.4152 | 0.5414 |
| 7 |  | 0.4422 | 0.4374 | 0.3527 | 0.3370 | 0.4383 | 0.8648 | 1.0992 | 0.6924 | 0.6634 | 0.6326 | 0.6772 | 0.9061 |
| +gp | F | 0.4422 . | 0.4374 | 0.3527 . | 0.3370 | ${ }^{0.4383}$ | 0.8648 | 1.0992 | ${ }^{0.6924}{ }_{\text {r }}$ | ${ }^{0.6634}$ | ${ }^{0.6326}$ | ${ }^{0.6772}$ | 0.9061 |
| FBAR 4-7 |  | 0.4411 | 0.4363 | 0.3520 | 0.3362 | 0.4372 | 0.8614 | 0.6895 | 0.5763 | 0.6537 | 0.5028 | 0.5280 | 0.5548 |
| Age/Year |  | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| 2 |  | 0.0247 | 0.0717 | 0.0256 | 0.1044 | 0.0258 | 0.0621 | 0.0272 | 0.0569 | 0.0693 | 0.1626 | 0.0903 | 0.2112 |
| 3 |  | 0.2952 | 0.2340 | 0.3452 | 0.4165 | 0.3090 | 0.2043 | 0.2397 | 0.2829 | 0.2887 | 0.5809 | 0.4866 | 0.4572 |
| 4 |  | 0.4283 | 0.5242 | 0.4723 | 0.6183 | 0.4662 | 0.3446 | 0.2390 | 0.3169 | 0.5238 | 0.4853 | 0.4471 | 0.6987 |
| 5 |  | 0.4721 | 0.5225 | 0.4959 | 0.5096 | 0.7134 | 0.5421 | 0.2442 | 0.2689 | 0.4636 | 0.4666 | 0.3051 | 0.6809 |
| 6 |  | 0.5200 | 0.5558 | 0.5846 | 0.4961 | 0.3936 | 0.7278 | 0.5121 | 0.3017 | 0.3192 | 0.3098 | 0.2895 | 0.4097 |
| 7 |  | 0.5468 | 0.4187 | 0.5347 | 0.7495 | 0.4032 | 0.2616 | 0.8314 | 0.5024 | 0.2148 | 0.1710 | 0.2037 | 0.4158 |
| +gp | F | ${ }^{0.5468}$ | ${ }^{0.4187}$ | 0.5347 . | 0.7495 | ${ }^{0.4032}$ | ${ }^{0.2616}$ | $0.8314{ }_{\text {\% }}$ | $0.5024{ }_{\text {F }}$ | 0.2148 | 0.1710 | 0.2037 . | 0.4158 |
| FBAR 4-7 |  | 0.4918 | 0.5053 | 0.5219 | 0.5934 | 0.4941 | 0.4690 | 0.4567 | 0.3475 | 0.3804 | 0.3582 | 0.3114 | 0.5513 |
| Age/Year |  | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| 2 |  | 0.0930 | 0.1009 | 0.0543 | 0.0433 | 0.0144 | 0.0284 | 0.0205 | 0.0381 | 0.0154 | 0.0088 | 0.0002 | 0.0014 |
| 3 |  | 0.3928 | 0.4203 | 0.2773 | 0.2993 | 0.1972 | 0.1793 | 0.3250 | 0.1428 | 0.0961 | 0.0527 | 0.0115 | 0.0108 |
| 4 |  | 0.5028 | 0.3230 | 0.3695 | 0.4054 | 0.2449 | 0.3386 | 0.3571 | 0.2422 | 0.1069 | 0.0881 | 0.0363 | 0.0142 |
| 5 |  | 0.4507 | 0.3150 | 0.2701 | 0.4761 | 0.3480 | 0.2632 | 0.3634 | 0.1764 | 0.1624 | 0.0656 | 0.0383 | 0.0295 |
| 6 |  | 0.5949 | 0.3094 | 0.3060 | 0.2633 | 0.3571 | 0.2901 | 0.2480 | 0.1484 | 0.0797 | 0.0939 | 0.0246 | 0.0162 |
| 7 |  | 0.3133 | 0.3955 | 0.3014 | 0.3872 | 0.2005 | 0.5003 | 0.2228 | 0.0956 | 0.1232 | 0.0521 | 0.0447 | 0.0174 |
| +gp |  | 0.3133 | 0.3955 | 0.3014 | 0.3872 | 0.2005 | 0.5003 | 0.2228 | 0.0956 | 0.1232 | 0.0521 | 0.0447 | 0.0174 |
| FBAR 4-7 | 「 | $0.4654^{\prime \prime}$ | 0.3357 | $0.3118^{\text {² }}$ | $0.3830^{\circ}$ | $0.287{ }^{\text {² }}$ | $0.3481^{\prime \prime}$ | $0.2978{ }^{\text {² }}$ | $0.1656^{\prime \prime}$ | $0.1181^{\text {² }}$ | $0.0749^{\prime \prime}$ | $0.0360^{\prime \prime}$ | 0.0193 |
| Age/Year |  | 2018 FB | R 16-18 |  |  |  |  |  |  |  |  |  |  |
| 2 |  | $0.0011^{\prime \prime}$ | 0.0009 |  |  |  |  |  |  |  |  |  |  |
| 3 |  | $0.0165^{\prime \prime}$ | 0.0130 |  |  |  |  |  |  |  |  |  |  |
| 4 |  | $0.0135^{\prime \prime}$ | 0.0213 |  |  |  |  |  |  |  |  |  |  |
| 5 |  | $0.0114^{\text {² }}$ | 0.0264 |  |  |  |  |  |  |  |  |  |  |
| 6 |  | $0.0195^{7}$ | 0.0201 |  |  |  |  |  |  |  |  |  |  |
| 7 |  | $0.010{ }^{\prime \prime}$ | 0.0242 |  |  |  |  |  |  |  |  |  |  |
| +gp |  | $0.010{ }^{\text {" }}$ | 0.0242 |  |  |  |  |  |  |  |  |  |  |
| FBAR 4-7 | * | 0.0138 |  |  |  |  |  |  |  |  |  |  |  |

Table 29.11. Sol.27.7a - Stock numbers-at-age (start of year, in thousands).

| Age/Year |  | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 |  | 3695 | 10177 | 3186 | 13133 | 5870 | 6680 | 3857 | 15772 | 9041 | 8848 | 5071 | 4498 | 2462 |
| 3 |  | 8349 | 3316 | 9101 | 2853 | 11533 | 5288 | 5795 | 3462 | 14061 | 8118 | 7904 | 4410 | 4003 |
| 4 |  | 4145 | 6703 | 2588 | 7594 | 2236 | 9588 | 4087 | 4887 | 2737 | 11812 | 6369 | 6258 | 3439 |
| 5 |  | 1368 | 2791 | 4071 | 1647 | 4784 | 1475 | 6406 | 2432 | 3193 | 1859 | 7423 | 3890 | 4075 |
| 6 |  | 4389 | 794 | 1451 | 2221 | 960 | 2699 | 822 | 3581 | 1464 | 1930 | 894 | 3809 | 2112 |
| 7 |  | 939 | 2585 | 498 | 802 | 1234 | 504 | 1642 | 509 | 2227 | 905 | 1140 | 313 | 1886 |
| +gp | V | 8212 | 5534 | 4321. | 3418 | 2828 | 3220 | 2221 | 2192 | 2042 | 1713 | ${ }^{2536}$ | 2365 | 1163 |
| TOTAL |  | 31097 | 31900 | 25214 | 31667 | 29446 | 29455 | 24830 | 32835 | 34765 | 35186 | 31338 | 25543 | 19140 |
| Age/Year |  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| 2 |  | 5559 | 15479 | 16250 | 23752 | 3462 | 3498 | 4374 | 5564 | 12686 | 4955 | 6191 | 5250 | 2006 |
| 3 |  | 2220 | 4995 | 13387 | 14557 | 21357 | 2952 | 3135 | 3788 | 4497 | 10227 | 4138 | 5523 | 4635 |
| 4 |  | 3293 | 1852 | 3772 | 10591 | 10000 | 16141 | 2250 | 2103 | 2300 | 2862 | 6941 | 3228 | 3720 |
| 5 |  | 1931 | 2031 | 1314 | 2381 | 6287 | 5145 | 10089 | 1298 | 964 | 1281 | 1717 | 4412 | 1903 |
| 6 |  | 2453 | 1026 | 1257 | 863 | 1240 | 2660 | 2658 | 5344 | 606 | 587 | 631 | 1022 | 2490 |
| 7 |  | 1234 | 1499 | 602 | 820 | 557 | 320 | 1162 | 1284 | 2629 | 333 | 351 | 332 | 550 |
| +gp | $\nabla$ | 2095 | 1963 | 2705 | 2501. | 1456 | 805 | 533 | 1005 | 1023 | 1972 | 1200 | 1302 | 1071 |
| TOTAL |  | 18784 | 28844 | 39288 | 55464 | 44360 | 31520 | 24200 | 20385 | 24705 | 22217 | 21170 | 21070 | 16375 |
| Age/Year |  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| 2 |  | 2499 | 8383 | 6899 | 5256 | 6968 | 4560 | 2332 | 3059 | 3644 | 2959 | 1321 | 1871 | 1962 |
| 3 |  | 1689 | 2204 | 6833 | 6083 | 4470 | 6136 | 3898 | 1969 | 2352 | 3013 | 2167 | 1090 | 1531 |
| 4 |  | 3319 | 1082 | 1315 | 4539 | 4487 | 3182 | 4184 | 2643 | 997 | 1308 | 1726 | 1324 | 648 |
| 5 |  | 1993 | 1872 | 528 | 746 | 2910 | 3197 | 2097 | 2242 | 1472 | 577 | 589 | 945 | 867 |
| 6 |  | 1021 | 1098 | 1018 | 234 | 393 | 2063 | 2211 | 1194 | 1272 | 982 | 264 | 339 | 624 |
| 7 |  | 1292 | 515 | 605 | 621 | 102 | 213 | 1380 | 1454 | 792 | 862 | 590 | 132 | 225 |
| +gp | $\cdots$ | 877 | 1053 | 1335 | 1057 | 476 | 721. | 1176 | 2300 | 1027. | 1132 | 1477 | 955 | 843 |
| TOTAL |  | 12690 | 16208 | 18533 | 18538 | 19806 | 20071 | 17278 | 14860 | 11557 | 10832 | 8134 | 6655 | 6700 |
| Age/Year |  | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | GMST 70 |  |  |
| 2 |  | 2278 | 1615 | 638 | 883 | 648 | 831 | 1807 | 4422 | 1553 | 3670 | 4087 |  |  |
| 3 |  | 1681 | 1974 | 1440 | 561 | 783 | 564 | 741 | 1621 | 4000 | 1403 |  |  |  |
| 4 |  | 1050 | 1128 | 1467 | 1089 | 367 | 614 | 464 | 636 | 1450 | 3580 G | ST 2009-2 |  |  |
| 5 |  | 405 | 633 | 799 | 946 | 690 | 261 | 499 | 384 | 555 | 1293 | 1339 |  |  |
| 6 |  | 599 | 228 | 405 | 556 | 595 | 523 | 200 | 423 | 335 | 487 |  |  |  |
| 7 |  | 416 | 417 | 144 | 274 | 392 | 464 | 437 | 165 | 374 | 298 |  |  |  |
| +gp | $\nabla$ | 528 | 912 | 324 | 604 | 945 | 509 | 1539 | 778 | 809 | 993 |  |  |  |
| TOTAL |  | 6957 | 6907 | 5216 | 4913 | 4420 | 3766 | 5687 | 8428 | 9074 | 11725 |  |  |  |

Table 29.12. Sol.27.7a - Summary.

|  | RECRUITS | SSB | BIOMASS | LANDINGS | FBAR 4-7 | YIELD/SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 2 |  |  |  |  |  |
| 1970 | 3695 | 6436 | 1785 | 1785 | 0.39 | 0.28 |
| 1971 | 10177 | 6222 | 1882 | 1882 | 0.44 | 0.30 |
| 1972 | 3186 | 5011 | 1450 | 1450 | 0.45 | 0.29 |
| 1973 | 13133 | 5123 | 1428 | 1428 | 0.43 | 0.28 |
| 1974 | 5870 | 5068 | 1307 | 1307 | 0.44 | 0.26 |
| 1975 | 6680 | 5359 | 1441 | 1441 | 0.40 | 0.27 |
| 1976 | 3857 | 4888 | 1463 | 1463 | 0.43 | 0.30 |
| 1977 | 15772 | 4490 | 1147 | 1147 | 0.37 | 0.26 |
| 1978 | 9041 | 5092 | 1106 | 1106 | 0.36 | 0.22 |
| 1979 | 8848 | 5685 | 1614 | 1614 | 0.47 | 0.28 |
| 1980 | 5071 | 5514 | 1941 | 1941 | 0.64 | 0.35 |
| 1981 | 4498 | 5166 | 1667 | 1667 | 0.48 | 0.32 |
| 1982 | 2462 | 4331 | 1338 | 1338 | 0.44 | 0.31 |
| 1983 | 5559 | 4096 | 1169 | 1169 | 0.44 | 0.29 |
| 1984 | 15479 | 4604 | 1058 | 1058 | 0.35 | 0.23 |
| 1985 | 16250 | 5642 | 1146 | 1146 | 0.34 | 0.20 |
| 1986 | 23752 | 6956 | 1995 | 1995 | 0.44 | 0.29 |
| 1987 | 3462 | 7165 | 2808 | 2808 | 0.86 | 0.39 |
| 1988 | 3498 | 5519 | 1999 | 1999 | 0.69 | 0.36 |
| 1989 | 4374 | 4633 | 1833 | 1833 | 0.58 | 0.40 |
| 1990 | 5564 | 3628 | 1583 | 1583 | 0.65 | 0.44 |
| 1991 | 12686 | 3189 | 1212 | 1212 | 0.50 | 0.38 |
| 1992 | 4955 | 3452 | 1259 | 1259 | 0.53 | 0.36 |
| 1993 | 6191 | 3231 | 1023 | 1023 | 0.55 | 0.32 |
| 1994 | 5250 | 4051 | 1374 | 1374 | 0.49 | 0.34 |
| 1995 | 2006 | 3521 | 1266 | 1266 | 0.51 | 0.36 |
| 1996 | 2499 | 2712 | 1002 | 1002 | 0.52 | 0.37 |
| 1997 | 8383 | 2500 | 1003 | 1003 | 0.59 | 0.40 |
| 1998 | 6899 | 3027 | 911 | 911 | 0.49 | 0.30 |
| 1999 | 5256 | 3320 | 863 | 863 | 0.47 | 0.26 |
| 2000 | 6968 | 3132 | 818 | 818 | 0.46 | 0.26 |
| 2001 | 4560 | 3565 | 1053 | 1053 | 0.35 | 0.30 |
| 2002 | 2332 | 3578 | 1090 | 1090 | 0.38 | 0.30 |
| 2003 | 3059 | 3213 | 1014 | 1014 | 0.36 | 0.32 |
| 2004 | 3644 | 2299 | 709 | 709 | 0.31 | 0.31 |
| 2005 | 2959 | 2076 | 855 | 855 | 0.55 | 0.41 |
| 2006 | 1321 | 1640 | 569 | 569 | 0.47 | 0.35 |


|  | RECRUITS | SSB | BIOMASS | LANDINGS | FBAR 4-7 | YIELD/SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age 2 |  |  |  |  |  |  |
| 2007 | 1871 | 1398 | 492 | 492 | 0.34 | 0.35 |
| 2008 | 1962 | 1334 | 332 | 332 | 0.31 | 0.25 |
| 2009 | 2278 | 1081 | 325 | 325 | 0.38 | 0.30 |
| 2010 | 1615 | 1204 | 277 | 277 | 0.29 | 0.23 |
| 2011 | 638 | 1103 | 330 | 330 | 0.35 | 0.30 |
| 2012 | 883 | 1166 | 298 | 298 | 0.30 | 0.26 |
| 2013 | 648 | 1099 | 148 | 148 | 0.17 | 0.13 |
| 2014 | 831 | 883 | 99 | 99 | 0.12 | 0.11 |
| 2015 | 1807 | 1316 | 76 | 76 | 0.07 | 0.06 |
| 2016 | 4422 | 1247 | 35 | 35 | 0.04 | 0.03 |
| 2017 | 1553 | 1891 | 34 | 34 | 0.02 | 0.02 |
| 2018 | 3670 | 2627 | 36 | 36 | 0.01 | 0.01 |
| Arith. Mean | 5538 | 3581 |  | 1054 | 0.41 | 0.28 |
| Units | (Thousands) | (Tonnes) |  | (Tonnes) |  |  |

Table 29.13. Sole in 7.a.

| Input for catch forecast and Fmsy analysis |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Input: | F 2019: TAC constraint for 2019 (414 t) |  |  |  |  |  |  |  |
|  | F 2020-2021: mean 16-18 |  |  |  |  |  |  |  |
| Catch and stock weights are mean 16-18 |  |  |  |  |  |  |  |  |
| Recruits age 2 in 2019-2021 GM(09-17) |  |  |  |  |  |  |  |  |
| 2019 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 1339 | 0.1 | 0.38 | 0 | 0 | 0.142667 | 0.0053 | 0.183333 |
| 3 | 3317 | 0.1 | 0.71 | 0 | 0 | 0.205667 | 0.07459 | 0.234 |
| 4 | 1249 | 0.1 | 0.97 | 0 | 0 | 0.261333 | 0.12291 | 0.300667 |
| 5 | 3196 | 0.1 | 0.98 | 0 | 0 | 0.336333 | 0.15195 | 0.377 |
| 6 | 1157 | 0.1 | 1 | 0 | 0 | 0.395333 | 0.11586 | 0.434333 |
| 7 | 432 | 0.1 | 1 | 0 | 0 | 0.430667 | 0.13952 | 0.464 |
| +gp | 1156 | 0.1 | 1 | 0 | 0 | 0.44038 | 0.1395 | 0.454563 |
| fbar 4-7 |  |  |  |  |  |  | 0.13256 |  |
| 2020 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 1339 | 0.1 | 0.38 | 0 | 0 | 0.142667 | 0.0009 | 0.183333 |
| 3 | 1205 | 0.1 | 0.71 | 0 | 0 | 0.205667 | 0.0130 | 0.234 |
| 4 | 2785 | 0.1 | 0.97 | 0 | 0 | 0.261333 | 0.0213 | 0.300667 |
| 5 | 999 | 0.1 | 0.98 | 0 | 0 | 0.336333 | 0.0264 | 0.377 |
| 6 | 2484 | 0.1 | 1 | 0 | 0 | 0.395333 | 0.0201 | 0.434333 |
| 7 | 932 | 0.1 | 1 | 0 | 0 | 0.430667 | 0.0242 | 0.464 |
| +gp | 1250 | 0.1 | 1 | 0 | 0 | 0.44038 | 0.0242 | 0.454563 |
| fbar 4-7 |  |  |  |  |  |  | 0.023024 |  |
| 2021 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 1339 | 0.1 | 0.38 | 0 | 0 | 0.142667 | 0.0009 | 0.183333 |
| 3 | 1210 | 0.1 | 0.71 | 0 | 0 | 0.205667 | 0.0130 | 0.234 |
| 4 | 1076 | 0.1 | 0.97 | 0 | 0 | 0.261333 | 0.0213 | 0.300667 |
|  | 2467 | 0.1 | 0.98 | 0 | 0 | 0.336333 | 0.0264 | 0.377 |
| 6 | 881 | 0.1 | 1 | 0 | 0 | 0.395333 | 0.0201 | 0.434333 |
| 7 | 2203 | 0.1 | 1 | 0 | 0 | 0.430667 | 0.0242 | 0.464 |
| +gp | 1927 | 0.1 | 1 | 0 | 0 | 0.44038 | 0.0242 | 0.454563 |
| fbar 4-7 |  |  |  |  |  |  | 0.023024 |  |

Table 29.14. Sol.27.7a - Management option table.
F 2019: TAC constraint for 2019 (414 t)
F 2020-2021: mean 16-18
Catch and stock weights are mean 16-18
Recruits age 2 in 2019-2021 GM (09-17)
Fbar age range: 4-7

| 2019 <br> Biomass | SSB | FMult | FBar | Landings |
| ---: | ---: | ---: | ---: | ---: |
| 3427 | 3079 | 5.757593 | 0.1330 | 400 |


| 2020 |  |  |  | 2021 |
| :---: | :---: | :---: | :---: | :---: |
| SSB | FMult | FBar | Landings | SSB |
| 3218 | 0.0000 | 0 | 0 | 3552 |
| 3218 | 0.1000 | 0.0023 | 7 | 3545 |
| 3218 | 0.2000 | 0.0046 | 15 | 3537 |
| 3218 | 0.3000 | 0.00691 | 22 | 3530 |
| 3218 | 0.4000 | 0.00921 | 29 | 3523 |
| 3218 | 0.5000 | 0.01151 | 37 | 3516 |
| 3218 | 0.6000 | 0.01381 | 44 | 3509 |
| 3218 | 0.7000 | 0.01612 | 51 | 3502 |
| 3218 | 0.8000 | 0.01842 | 59 | 3495 |
| 3218 | 0.9000 | 0.02072 | 66 | 3488 |
| 3218 | 1.0000 | 0.02302 | 73 | 3481 |
| 3218 | 1.1000 | 0.02533 | 80 | 3474 |
| 3218 | 1.2000 | 0.02763 | 88 | 3467 |
| 3218 | 1.3000 | 0.02993 | 95 | 3460 |
| 3218 | 1.4000 | 0.03223 | 102 | 3453 |
| 3218 | 1.5000 | 0.03454 | 109 | 3446 |
| 3218 | 1.6000 | 0.03684 | 116 | 3439 |
| 3218 | 1.7000 | 0.03914 | 123 | 3432 |
| 3218 | 1.8000 | 0.04144 | 130 | 3426 |
| 3218 | 1.9000 | 0.04375 | 138 | 3419 |
| 3218 | 2.0000 | 0.04605 | 145 | 3412 |

Input units are thousands and kg - output in tonnes

| 2020 |  |  | 2021 | 2021-2020 | 2020-2019 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FMult | Landings | FBar | SSB | SSB change | TAC change | Basis |
| 7.987 | 542 | 0.184 | 3028 | -5.90 | 36.00 | MSY approach |
| 8.6867 | 585 | 0.200 | 2987 | -7.20 | 46.00 | Fmsy |
| 6.3895 | 441 | 0.147 | 3126 | -2.90 | 10.33 | MSY approach upper |
| 9.5844 | 640 | 0.221 | 2934 | -8.90 | 60.00 | MSY approach lower |
| 5.76 | 400 | 0.133 | 3165 | -1.70 | 0.00 | TACstable |
| 6.6866 | 460 | 0.154 | 3107 | -3.50 | 15.00 | TACplus15 |
| 4.8467 | 340 | 0.112 | 3223 | 0.10 | -15.00 | TACminus15 |
| 9.121 | 612 | 0.210 | 2961 | -8.00 | 53.00 | Fpa |
| 12.596 | 815 | 0.290 | 2766 | -14.10 | 104.00 | Flim |
| 5.7575 | 400 | 0.133 | 3165 | -1.70 | 0.00 | F2019 |
| 0.7292 | 53 | 0.0168 | 3500 | 8.70 | -87.00 | Btrigger |
| 17.783 | 1091 | 0.409 | 2500 | -22.30 | 173.00 | Blim |

Table 29.15. Sol.27.7a - Detailed results.

| F 2019: TAC constraint for 2019 (414 t) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| F 2020-2021: mean 16-18 |  |  |  |  |  |  |  |
| Catch and stock weights are mean 16-18 |  |  |  |  |  |  |  |
| Recruits age 2 in 2019-2021 GM (09-17) |  |  |  |  |  |  |  |
| Fbar age range: 4-7 |  |  |  |  |  |  |  |
| Year: | 2019 | F multiplier: | 5.97312 | Fbar: | 0.13752 |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| 2 | 0.00530 | 7 | 1 | 1339 | 191 | 509 | 73 |
| 3 | 0.07459 | 227 | 53 | 3317 | 682 | 2355 | 484 |
| 4 | 0.12291 | 138 | 41 | 1249 | 326 | 1211 | 317 |
| 5 | 0.15195 | 429 | 162 | 3196 | 1075 | 3132 | 1053 |
| 6 | 0.11586 | 121 | 52 | 1157 | 457 | 1157 | 457 |
| 7 | 0.13952 | 54 | 25 | 432 | 186 | 432 | 186 |
| 8 | 0.13952 | 143 | 65 | 1156 | 509 | 1156 | 509 |
| Total | 0.13256 | 1119 | 400 | 11846 | 3427 | 9952 | 3079 |
| Year: | 2020 | F multiplier: | 1.00000 | Fbar: | 0.02302 |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| 2 | 0.00092 | 1 | 0 | 1339 | 191 | 509 | 73 |
| 3 | 0.01295 | 15 | 3 | 1205 | 248 | 856 | 176 |
| 4 | 0.02135 | 56 | 17 | 2785 | 728 | 2702 | 706 |
| 5 | 0.02639 | 25 | 9 | 999 | 336 | 979 | 329 |
| 6 | 0.02012 | 47 | 20 | 2484 | 982 | 2484 | 982 |
| 7 | 0.02423 | 21 | 10 | 932 | 402 | 932 | 402 |
| 8 | 0.02423 | 28 | 13 | 1250 | 550 | 1250 | 550 |
| Total | 0.02302 | 193 | 73 | 10994 | 3437 | 9712 | 3218 |
| Year: | 2021 | F multiplier: | 1.00000 | Fbar: | 0.02302 |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| 2 | 0.00092 | 1 | 0 | 1339 | 191 | 509 | 73 |
| 3 | 0.01295 | 15 | 3 | 1210 | 249 | 859 | 177 |
| 4 | 0.02135 | 22 | 7 | 1076 | 281 | 1044 | 273 |
| 5 | 0.02639 | 61 | 23 | 2467 | 830 | 2418 | 813 |
| 6 | 0.02012 | 17 | 7 | 881 | 348 | 881 | 348 |
| 7 | 0.02423 | 50 | 23 | 2203 | 949 | 2203 | 949 |
| 8 | 0.02423 | 44 | 20 | 1927 | 849 | 1927 | 849 |
| Total | 0.02302 | 210 | 84 | 11103 | 3697 | 9841 | 3481 |



Figure 29.1a. Sol.27.7a - Age composition of landings.


Figure 29.1b. Sol.27.7a - Age composition of landings.

## Standardized catch (L) proportion at age



Figure 29.2. Sol.27.7a - Standardized catch proportion.


Figure 29.3a. Sol.27.7a - BE Length distributions of discarded and retained fish from discard sampling studies (Beam trawl).


Figure 29.3b. Sol.27.7a - IR Length distributions of discarded and retained fish from discard sampling studies (Otter trawl).

Figure 29.4a Sole in 7.a - Effort series


Figure 29.4b Sole in 7.a - Relative effort series


Figure 29.4c Sole in 7.a - Relative LPUE series


Figure 29.5b Sole in 7.a. Relative effort series


Figure 29.5c Sole in 7.a. Relative LPUE series


Figure 29.6-Sol.27.7.a - Mean-standardised indices


Figure 29.7 - Sol.27.7a - Consistency plot UK(E\&W)-BTS-Q3 survey


Figure 29.8 - Sol.27.7a - LOG CATCHABILITY RESIDUAL PLOTS - Final XSA
Residuals


Figure 29.9-Sol.27.7a - Summary plots


Figure 29.10-Sol.27.7a - Retrospective XSA analysys (shinkage SE=1.5)


Figure 29.11-Sol.27.7a - comparison with last year's assessment




Figure 29.12-Sol.27.7a -
Year-class sources and contributions for the short-term forecast


Figure 29.13. Sol.27.7a - Stock-recruitment plot


## 32 Sole (Solea solea) in divisions 7.b and 7.c (West of Ireland)

## Type of assessment in 2019

No assessment was performed.

### 32.1 General

### 32.1.1 Stock identity

Sole in 7.b are mainly caught by Irish vessels on sandy grounds in coastal areas. Sole catches in 7.c are negligible. In 7.b there are two distinct areas where sole are caught: an area around Galway Bay and an area in the north of $7 . \mathrm{b}$ which extends into $6 . a$ (the Stags and Broadhaven Ground). The landings and lpue of sole in 7.bc appear to have been more or less stable since the start of the logbooks' time-series in 1995 (WD1, WGCSE 2009). It is not known how much exchange there is between sole on the Aran Grounds and those on the Stags Ground.

### 32.1.2 Data

The time-series of official landings is presented in Table 30.1 and Figure 30.1.
The time-series of otter-trawl landings effort and lpue since 1995 are shown in Figure 30.2. Lpue shows no trend over the time-series but has fluctuated more in recent years.
Sampling is carried out in Ireland but numbers of samples varies over time due to the low landings levels and varying encounter probability and is not sufficient to generate a time-series of annual length or age distributions. Sampling in 2018 comprised of 12 length samples, totalling 938 fish measured, and eight discard trips. The length distribution for the eight discard trips is shown in Figure 30.3.

### 32.1.3 Historical stock development

No analytical assessment was performed.

Table 30.1. Landings of Sole in 7.bc as officially reported to ICES.

| Year | BEL | FRA | UK | IRL | OTH | TOT | Year | BEL | FRA | UK | IRL | OTH | TOT | Unalloc | WG est |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1908 | 0 | 0 | 1 | 37 | 0 | 38 | 1963 | 0 | 172 | 0 | 19 | 0 | 191 |  |  |
| 1909 | 0 | 0 | 0 | 32 | 0 | 32 | 1964 | 0 | 159 | 1 | 24 | 0 | 184 |  |  |
| 1910 | 0 | 0 | 0 | 28 | 0 | 28 | 1965 | 0 | 95 | 5 | 24 | 0 | 124 |  |  |
| 1911 | 0 | 0 | 1 | 22 | 0 | 23 | 1966 | 0 | 0 | 1 | 11 | 0 | 12 |  |  |
| 1912 | 0 | 0 | 1 | 22 | 0 | 23 | 1967 | 0 | 78 | 0 | 11 | 0 | 89 |  |  |
| 1913 | 0 | 0 | 1 | 25 | 0 | 26 | 1968 | 0 | 121 | 0 | 8 | 0 | 129 |  |  |
| 1914 | 0 | 0 | 1 | 43 | 0 | 44 | 1969 | 0 | 86 | 1 | 9 | 0 | 96 |  |  |
| 1915 | 0 | 0 | 1 | 12 | 0 | 13 | 1970 | 0 | 3 | 0 | 8 | 0 | 11 |  |  |
| 1916 | 0 | 0 | 0 | 14 | 0 | 14 | 1971 | 0 | 0 | 2 | 5 | 0 | 7 |  |  |
| 1917 | 0 | 0 | 0 | 6 | 0 | 6 | 1972 | 0 | 4 | 0 | 13 | 0 | 17 |  |  |
| 1918 | 0 | 0 | 0 | 7 | 0 | 7 | 1973 | 0 | 0 | 0 | 12 | 0 | 12 |  |  |
| 1919 | 0 | 0 | 0 | 6 | 0 | 6 | 1974 | 0 | 25 | 0 | 12 | 0 | 37 |  |  |
| 1920 | 0 | 0 | 9 | 5 | 0 | 14 | 1975 | 0 | 7 | 0 | 19 | 0 | 26 |  |  |
| 1921 | 0 | 0 | 10 | 9 | 0 | 19 | 1976 | 0 | 6 | 0 | 44 | 0 | 50 |  |  |
| 1922 | 0 | 0 | 4 | 9 | 0 | 13 | 1977 | 0 | 3 | 0 | 14 | 0 | 17 |  |  |
| 1923 | 0 | 0 | 2 | 10 | 0 | 12 | 1978 | 0 | 3 | 0 | 16 | 0 | 19 |  |  |
| 1924 | 0 | 0 | 15 | 64 | 0 | 79 | 1979 | 0 | 6 | 0 | 13 | 0 | 19 |  |  |
| 1925 | 0 | 0 | 11 | 18 | 0 | 29 | 1980 | 0 | 9 | 0 | 24 | 0 | 33 |  |  |
| 1926 | 0 | 7 | 10 | 18 | 0 | 35 | 1981 | 0 | 6 | 0 | 47 | 0 | 53 |  |  |
| 1927 | 0 | 47 | 11 | 19 | 0 | 77 | 1982 | 0 | 5 | 1 | 55 | 0 | 61 |  |  |
| 1928 | 0 | 49 | 8 | 16 | 0 | 73 | 1983 | 0 | 9 | 0 | 40 | 0 | 49 |  |  |
| 1929 | 0 | 74 | 11 | 18 | 0 | 103 | 1984 | 0 | 3 | 0 | 17 | 0 | 20 |  |  |
| 1930 | 0 | 52 | 5 | 22 | 0 | 79 | 1985 | 0 | 6 | 0 | 44 | 0 | 50 |  |  |
| 1931 | 0 | 82 | 9 | 29 | 0 | 120 | 1986 | 0 | 8 | 0 | 29 | 0 | 37 |  |  |
| 1932 | 0 | 122 | 10 | 27 | 0 | 159 | 1987 | 0 | 2 | 0 | 39 | 0 | 41 |  |  |
| 1933 | 0 | 411 | 10 | 10 | 0 | 431 | 1988 | 0 | 2 | 1 | 34 | 0 | 37 |  |  |
| 1934 | 0 | 217 | 10 | 13 | 0 | 240 | 1989 | 0 | 0 | 0 | 38 | 0 | 38 |  |  |
| 1935 | 0 | 40 | 7 | 11 | 0 | 58 | 1990 | 0 | 0 | 0 | 41 | 0 | 41 |  |  |


| Year | BEL | FRA | UK | IRL | OTH | TOT | Year | BEL | FRA | UK | IRL | OTH | TOT | Unalloc | WG est |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1936 | 0 | 43 | 20 | 9 | 0 | 72 | 1991 | 0 | 5 | 0 | 46 | 0 | 51 |  |  |
| 1937 | 0 | 32 | 25 | 14 | 0 | 71 | 1992 | 0 | 2 | 0 | 43 | 0 | 45 |  |  |
| 1938 | 0 | 44 | 21 | 7 | 0 | 72 | 1993 | 0 | 1 | 0 | 59 | 0 | 60 | 0 | 60 |
| 1939 | 0 | 0 | 0 | 13 | 0 | 13 | 1994 | 0 | 1 | 0 | 60 | 0 | 61 | 9 | 70 |
| 1940 | 0 | 0 | 0 | 19 | 0 | 19 | 1995 | 0 | 2 | 0 | 59 | 0 | 61 | -2 | 59 |
| 1941 | 0 | 0 | 0 | 14 | 0 | 14 | 1996 | 0 | 2 | 0 | 52 | 0 | 54 | 3 | 57 |
| 1942 | 0 | 0 | 0 | 8 | 0 | 8 | 1997 | 0 | 3 | 1 | 51 | 0 | 55 | 0 | 55 |
| 1943 | 0 | 0 | 0 | 11 | 0 | 11 | 1998 | 0 | 0 | 0 | 49 | 0 | 49 | 17 | 66 |
| 1944 | 0 | 0 | 0 | 16 | 0 | 16 | 1999 | 0 | 0 | 0 | 68 | 0 | 68 | 4 | 72 |
| 1945 | 0 | 0 | 0 | 20 | 0 | 20 | 2000 | 0 | 12 | 0 | 65 | 0 | 77 | -9 | 68 |
| 1946 | 0 | 0 | 12 | 10 | 0 | 22 | 2001 | 0 | 7 | 0 | 53 | 0 | 60 | 0 | 60 |
| 1947 | 15 | 0 | 6 | 8 | 0 | 29 | 2002 | 0 | 14 | 0 | 50 | 0 | 64 | -3 | 61 |
| 1948 | 0 | 0 | 11 | 14 | 0 | 25 | 2003 | 0 | 19 | 0 | 50 | 0 | 69 | -5 | 64 |
| 1949 | 0 | 41 | 12 | 12 | 0 | 65 | 2004 | 0 | 18 | 0 | 49 | 0 | 67 | 2 | 69 |
| 1950 | 0 | 24 | 9 | 6 | 0 | 39 | 2005 | 0 | 7 | 0 | 38 | 0 | 45 | -1 | 44 |
| 1951 | 0 | 27 | 7 | 6 | 0 | 40 | 2006 | 0 | 12 | 0 | 31 | 0 | 43 | 0 | 43 |
| 1952 | 0 | 40 | 2 | 6 | 0 | 48 | 2007 | 0 | 7 | 0 | 34 | 0 | 41 | 1 | 42 |
| 1953 | 0 | 99 | 2 | 4 | 0 | 105 | 2008 | 0 | 6 | 0 | 31 | 0 | 37 | 3 | 40 |
| 1954 | 0 | 116 | 1 | 7 | 0 | 124 | 2009 | 0 | 5 | 0 | 46 | 0 | 51 | 0 | 51 |
| 1955 | 0 | 66 | 1 | 9 | 0 | 76 | 2010 | 0 | 8 | 0 | 35 | 0 | 43 | 0 | 43 |
| 1956 | 0 | 161 | 1 | 6 | 0 | 168 | 2011 | 0 | 5 | 0 | 22 | 0 | 27 | -5 | 22 |
| 1957 | 0 | 94 | 1 | 4 | 0 | 99 | 2012 | 0 | 7 | 0 | 38 | 0 | 45 | -2 | 43 |
| 1958 | 0 | 163 | 2 | 6 | 0 | 171 | 2013 | 0 | 3 | 0 | 30 | 0 | 33 | 0 | 33 |
| 1959 | 0 | 327 | 1 | 8 | 0 | 336 | 2014 | 0 | 3 | 0 | 23 | 0 | 26 | 1 | 27 |
| 1960 | 0 | 80 | 1 | 9 | 0 | 90 | 2015 | 0 | 3 | 0 | 31 | 0 | 34 | 0 | 34 |
| 1961 | 0 | 110 | 1 | 12 | 0 | 123 | 2016 | 0 | 6 | 0 | 36 | 0 | 42 | 0 | 42 |
| 1962 | 0 | 100 | 0 | 8 | 0 | 108 | 2017 | 0 | 5 | 0 | 22 | 0 | 27 | 0 | 27 |
|  |  |  |  |  |  |  | 2018 | 0 | 5 | 1 | 22 | 0 | 28 | 0 | 18 |



Figure 30.1. Landings of Sole in 7.bc as officially reported to ICES (1908-2017).


Figure 30.2. Sole in 7.b Irish otter trawl landings effort and lpue since 1995.


Figure 30.3. Estimated age distribution of sole 7.bc catches in 2018 vessels, at sea sampling on OTB based on Irish sampling.

## 33 Sole (Solea solea) in Division 7.e (western English Channel)

## Type of assessment in 2018

Last year's assessment report is available at:
http://www.ices.dk/sites/pub/Publication\ Reports/Expert\ Group\ Report/acom/2018/WGCSE/35 Sole 7e 2018.pdf

ICES advice applicable to 2019
Last year's advice is available at:
http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2018/2018/sol.27.7e.pdf

### 33.1 General

## Stock description and management units

The TAC specified for ICES Division 7.e is consistent with the assessment area.
Official national landings data as reported to ICES and the landings estimates as used by the Working Group are given in Table 31.1.

Official landings in 2018 were 1074 t , a 10.7\% undershoot of the TAC (1202 t).
The TAC and the national quotas by country for 2018
\(\left.$$
\begin{array}{lcl}\hline \text { Species: } & \begin{array}{l}\text { Sole } \\
\text { Solea solea }\end{array} & \text { Zone }\end{array}
$$ \begin{array}{l}7.e <br>

(Sol/07E.)\end{array}\right]\)| Belgium | 42 |  |
| :--- | :---: | :--- |
| France | 453 | Analytical TAC |
| United Kingdom | 707 | Article 7(2) of this Regulation applies |
| Union | 1202 |  |
| TAC |  |  |

Article 7(2): " 2 . The stocks of non-target species within safe biological limits referred to in Article 15(8) of Regulation (EU) No 1380/2013 are identified in Annex I to this Regulation for the purposes of the derogation from the obligation to count catches against the relevant quotas provided for in that Article."
(Source: Council Regulation (EU) 2018/120, ANNEX IA, EU, 2018b)

The TAC and the national quotas by country for 2019
\(\left.$$
\begin{array}{lcl}\hline \text { Species: } & \begin{array}{l}\text { Sole } \\
\text { Solea solea }\end{array} & \text { Zone }\end{array}
$$ \begin{array}{l}7.e <br>

(Sol/07E.)\end{array}\right]\)| Belgium | 44 |  |
| :--- | :---: | :--- |
| France | 468 | Analytical TAC |
| United Kingdom | 730 | Article 7(2) of this Regulation applies |
| Union | 1242 |  |
| TAC |  |  |

Article 7(2): " 2 . The stocks of non-target species within safe biological limits referred to in Article 15(8) of Regulation (EU) No 1380/2013 are identified in Annex I to this Regulation for the purposes of the derogation from the obligation to count catches against the relevant quotas provided for in that Article."
(Source: Council Regulation (EU) 2019/124, EU, 2019a)
Maximum number of days a vessel may be present within the area by category of regulated gear per year for 2018 and 2019

| Regulated gear | Maximum number of days |  |
| :--- | :--- | :--- |
| Beam trawls of mesh size $\geq 80 \mathrm{~mm}$ | BE | 176 |
|  | FR | 188 |
| Static nets with mesh size $\leq 220 \mathrm{~mm}$ | BE | 222 |
|  | FR | 176 |

(Source: Council Regulation (EU) 2018/120, ANNEX IIC, EU, 2018b and Council Regulation (EU) 2019/124, EU, 2019a)

## Landing obligation

As of 2019, the landing obligation fully applies to sole in 7.e.
The landing obligation was phased in between 2016-2019 (Commission Delegated Regulations (EU) 2015/2438, 2016/2375, 2018/46, EU, 2015, 2016, 2018a). During the phasing in, the landing obligation applied to all catches of sole in $7 . e$ with trammel and gillnets (gear codes GNS, GN, GND, GNC, GTN, GTR, GEN) and all beam trawls. However, a de minimis exemption applied, allowing up to $3 \%$ discards of annual catches for all trammel and gillnets and for beam trawls with mesh size of $80-199 \mathrm{~mm}$ with increased selectivity. In 2016, the first year of the application, the landing obligation applied only to vessels for which the total landings consisted of more than $10 \%$ sole during two reference years (2013 and 2014, Commission Delegated Regulation (EU) 2015/2438, EU, 2015). This threshold was tightened for 2017 and the landing obligation applied to vessels landing more than 5\% in the reference years 2014 and 2015 (Commission Delegated

Regulation (EU) 2016/2375, EU, 2016). Subsequently, this restriction was lifted altogether and for 2018 (Commission Delegated Regulation (EU) 2018/46, EU, 2018a), the landing obligation applied to all vessels using trammel and gillnets and beam trawls, as described above.

Given the low discards observed in the fishery the landing obligation is unlikely to have a significant impact on this stock or the advice.

### 33.2 Data

## InterCatch

International catch data are collated using the ICES InterCatch platform. For 2018, data for Belgium, France, Ireland, the Netherlands and the UK (England, Scotland and the Channel Islands Guernsey and Jersey) were uploaded into InterCatch (Figures 31.1 and 31.2). All submitted age samples are presented in Figures 31.8 and 31.9, and length samples in Figures 31.10 and 31.11. The raising procedure is described in the stock annex.

## Landings

Landings of sole in Division 7.e were below 500 t at the beginning of the time-series in the 1970s, increased and stayed around 1500 t in the 1980s and have been around 1000 t in the 1990s and 2000s (Table 31.1). The landings dropped in the late 2000s below 750 t and increased since 2015 to 1075 t in 2018.

Only the UK and France provided age-structured landings samples in InterCatch (Figures 31.8 and 31.9).

Total international landings numbers-at-age (Table 31.2 and Figure 31.5) and landings and stock weights-at-age (Tables 31.3 and 31.4 and Figure 31.6) as used in the assessment were derived in accordance with the procedures outlined in the stock annex.

The fleets for which age-distributions were submitted, accounted for $83.1 \%$ of the total international landings.

## Discards

Discards for this stock are low and not included in the assessment.
For 2017, discards data were provided by Belgium, France, Ireland (zero discards) and the UK (zero discards) for some fleets in InterCatch based on discard sampling. Age samples were provided by France for two fleets. As discard are considered to be low, discards were not raised to an international level or allocated with an age structure from sample data.
Discards data are only available from InterCatch for the years 2012-2018. In general, the discard rates are low (Figure 31.3). The discards provided in InterCatch accounted for $0.44 \%$ in 2016 and $0.26 \%$ in 2018 of the total catches. The three-year average (2016-2018) discard rate is $0.61 \%$. The reduction in the discard rate might potentially be linked to the introduction of the landing obligation in 2016.

The discard rate by fleet and country is shown in Figure 31.3 (shown are only discards submitted to InterCatch).

No discard information is included in the assessment given that it is currently not possible to provide discard estimates for the entire time-series. Nevertheless, excluding discard estimates from the assessment is unlikely to have any major impact on the perception of stock status given the minor scale of the problem.

## Revisions

No revision to previous years were submitted.

## Biological

Natural mortality was assumed to be constant over ages and years at 0.1 . The maturity ogive from divisions 7.f and 7.g was used in accordance with the procedures outlined in the stock annex and adopted in previous assessments.

In agreement with the stock annex stock and catch and stock weights-at-age were derived by fitting a 2 nd degree polynomial model to the raw landings weights-at-age extracted from InterCatch (Figure 31.7). For 2018 data, the youngest age for which data (catch numbers and weights) was provided was age 0 . In previous years, this data was only provided for ages 1 and older. For historical consistency and a better model fit, age 0 was removed from the smoothing process for catch and stock weights.

## Survey indices

Abundance estimates derived from the surveys as used in the assessment are given in Table 31.6 and shown in Figures 31.13, 31.14 and 31.15, and internal consistencies in Figures 31.16 and 31.17. In general, cohort tracking and internal consistency are better in the commercial tuning fleets and less pronounced in the scientific surveys.

## The UK-FSP survey

The UK Fisheries Science Partnership (UK-FSP) conducted another survey of sole and plaice abundance in the Western English Channel, 2018 being its 16th year (the first years is not used for sole due to data issues). The results indicate that sole continue to be widespread in the area and that a large number of cohorts contribute to the stock. The total cpue increased from 2010 until 2014, dropped subsequently until 2016 and increased again since then. The index is mainly driven by ages 3,4 and 5 . The internal consistency in the survey is good for ages $3+$. Some year and cohort effects are visible.

## The QISWBeam survey

Abundance estimates for the Quarter 1 South West Beam trawl (Q1SWBeam) survey started in 2006 and have been included in the assessment since 2014. The landings per unit of effort (lpue) numbers-at-age as well as aggregated over all ages are variable without particular trends or patterns. As for the UK-FSP, the index mainly consisted of fish age 3, 4 and 5 in 2018. Internal consistency is mediocre and no particular good cohort tracking is evident.

## Commercial fleets effort and Ipue

Two commercial tuning series from the UK are used (commercial beam trawl UK-CBT and commercial otter trawl UK-COT).

Effort for under 24 m UK beam trawlers in days fished steadily increased from 1992 reaching the highest levels on record in 2012, and stayed around this level until the end of the time-series (Figure 31.12). Currently the effort is well above the long-term average. In contrast, effort for over 24 m UK beam trawlers increased from 1992 to 2004 and then decreased to below the average of the time-series, reaching a minimum in 2013. Since then, the effort increased again slightly and is currently around the long-term average. When the effort of all UK beam trawl vessels is combined, the effort stayed almost constant since the early 2000s.

UK otter trawl (UK-COT) effort has been in continual decline since the early-1970s and was at the lowest levels on record in 2015. For 2016, this fleet reported zero effort and landings. This
could be explained by a shift in the size of fishing vessels to smaller vessels. Since 2017, a new database is being used for recording, but the data are not consistent with historical data and are therefore not used in the stock assessment.

Age-disaggregated commercial abundance indices for the UK-CBT-late (UK-CBT values from 2003 onwards) and UK-COT fleets as used in the assessment are given in Table 31.5 and plotted mean standardised by cohort and year in Figures 31.13 and 31.14.

## Information from the fishing industry

No comments were received regarding the assessment or management of this stock beyond the information from the UK fisheries-science partnership already formally included in the assessment process.

### 33.3 Stock assessment

Model used: Extended Survivors Analysis (XSA) as outlined in the stock annex by IBPWCFlat2 2015.

Software used: FLR - FLXSA.
Model options chosen: Data included in the assessment were identical to previous years.
Assessment input data characteristics: catch numbers-at-age excluding discards and with four tuning fleets (two fishery-independent surveys: UK-FSP and Q1SWBeam; and two commercial lpue time-series: UK-CBT-late and UK-COT).

## Data screening

Data screening procedures identified no major anomalies in the catch numbers-at-age, weights or tuning information used in the 2019 assessment.

The landings numbers-at-age 3 were exceptionally high in 2017 but returned to usual levels in 2018 (Figure 31.5). This anomaly was evident in age samples from the UK and from France and in various fleets (see WGCSE 2018 report), i.e. does not seem to be a sampling issue.
Tuning information consisted of four fleets: two UK commercial time-series (UK-CBT-late and UK-COT) and two UK standardised research surveys (UK-FSP and Q1SWBeam).

The UK commercial otter trawl fleet (UK-COT) reported zero effort in 2016 and therefore there is no lpue value for this fleet for 2016. Consequently, this tuning index only influences the assessment up to and including 2015.

Details of the derivation of the tuning fleets are presented in the stock annex, and the tuning information available for this assessment is shown in Table 31.6 and Figures 31.13, 31.14 and 31.15 .

## Final update assessment

The working group fitted the XSA model developed by WKFLAT 2012 (ICES, 2012) using the updated assessment settings agreed at IBPWCFlat2 (ICES, 2015).

The XSA assessment settings used at the last three working groups are shown in the table below, and more historical settings have been included in the stock annex.

|  | WGCSE 2017 | WGCSE 2018 | WGCSE 2019 |
| :---: | :---: | :---: | :---: |
| Assessment age range | 2-12+ | 2-12+ | 2-12+ |
| Fbar age range | $F(3-9)$ | F(3-9) | F(3-9) |
| Assessment method | XSA | XSA | XSA |
| Tuning Fleets: |  |  |  |
| Q1SWBeam | 2006-2016 | 2006-2017 | 2006-2018 |
|  | Ages 2-11 (non-offset) | Ages 2-11 (non-offset) | $\begin{aligned} & \text { Ages 2-11 } \\ & \text { (non-offset) } \end{aligned}$ |
| UK-FSP | 2004-2016 | 2004-2017 | 2004-2018 |
|  | Ages 2-11 | Ages 2-11 | Ages 2-11 |
| UK combined beam (late) | 2003-2016 | 2003-2017 | 2003-2018 |
|  | Ages 3-11 | Ages 3-11 | Ages 3-11 |
| UK otter trawl | 1988-2016 | 1988-2016 | 1988-2016 |
|  | Ages 3-11 | Ages 3-11 | Ages 3-11 |
| Time taper | Yes | Yes | Yes |
| Power model | Tricubic | Tricubic | Tricubic |
| Taper range | 15 years | 15 years | 15 years |
| P shrinkage | No | No | No |
| Q plateau age | 7 | 7 | 7 |
| F shrinkage S.E | 0.5 | 0.5 | 0.5 |
| Number of years | 3 | 3 | 3 |
| Number of ages | 5 | 5 | 5 |
| Fleet S.E. | 0.4 | 0.4 | 0.4 |

Figure 33.18 shows the results from the final XSA model fit, Figure 31.19 the residuals, Figure 31.20 a comparison of the current assessment with the assessment from the last two years, Figure 31.21 XSA survivor weightings for the last years and Figure 31.22 a five-year retrospective.

The WGCSE 2018 assessment showed a rescaling (downwards revision of SSB). A similar rescaling occurred for the current (WGCSE 2019) assessment, but in the opposite direction, i.e. the absolute levels are now again close to the ones from WGCSE 2016. This rescaling can most likely be attributed to the landings data from 2017, which had a high number of landings-at-age three, which returned again to normal in the 2018 data.

A Mohn's rho analysis was conducted based on the XSA stock assessment results, i.e. the last data year (2018) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (ages 3-9) | recruitment |
| :--- | :--- | ---: | :--- |
| Mohn's rho value | -0.00759231890293091 | 0.0406781290651195 | 0.00784258485351308 |

The Mohn's rho values for this assessment are very low and well below the threshold of $20 \%$ imposed by ICES for 2019 assessments, i.e. the current assessment indicates a high consistency.

XSA diagnostic tables, stock numbers-at-age and fishing mortalities-at-age for the final assessment are shown in Tables 31.7, 31.8, 31.9 and 31.10.

## State of the stock

Stock trends are shown in Table 31.10 and plotted in Figure 31.18. The stock is in a desirable state, both in term of spawning-stock biomass and fishing mortality.

SSB is estimated to have increased between 1972 and 1980 following successive strong recruitment events. Subsequently, SSB declined from 1980 to 1993 and remained relatively stable until 2008. After this period, SSB has been increasing and is currently well above MSY B ${ }_{\text {trigger }}$.

The base level of recruitment has remained relatively stable throughout the time-series, fluctuating without major temporal trend at around 4 million recruits. Recruitment variability has decreased since 1991. Recruitment has been at or above the long-term geometric mean in the last four years.

Fishing mortality was relatively stable at a low level between 1969 and 1978, after which it increased sharply until 1983 and fluctuated at a higher level before peaking briefly in 1989-1990. After a period of temporal variability, F decreased abruptly to below the Fmsy target of 0.29 in 2009 and has remained below this level ever since and stayed around $0.2-0.25$. Fishing mortality was estimated to be well below all reference points.

The age structure of sole in 7.e continues to be more extended than other sole stocks in European waters, implying low mortality rates, with the plus group at-age 12 containing a high proportion of the catches and including some individuals aged 33-38 in recent years.

### 33.4 Short-term projections

## Forecast assumptions

Figure 31.23 shows three different targets for the intermediate year ( $\mathrm{F}=\mathrm{F}_{2018}, \mathrm{~F}=\mathrm{F}_{2016-2018}$ and landings $=\mathrm{TAC}_{2019}$ ). F estimates in 2016-2018 fluctuated around 0.2 but are slightly higher in 2018. Consequently, $\mathrm{F}_{2018}(0.23)$ is used for the intermediate year. Landings were below the TAC in the last three years ( $10.7 \%$ below in 2018 ) and consequently, the selected F is an appropriate target for the intermediate year.

Weights-at-age were calculated as the average of three last three historical years.
Recruitment was forecast using a long-term geometric mean (1969-2018) due to temporal variability in the time-series and the lack of distinct periods of successive high or low recruitment in recent years.

The forecast was conducted with FLR's FLash using the output from the landings only XSA assessment. The resulting yield was obtained by adding discards to the landing with an average discard rate of the last three historical years (2016-2018, 0.61\%).
The complete input data for the short-term forecast are shown in Table 31.11.

## MSY forecast

Table 31.12 shows a detailed output of the forecast targeting Fmsy for the years 2019-2021 and Table 31.13 the year classes contributing to the yield in 2020 and SSB in 2021.

Figure 31.24 shows the results of the forecast for $\mathrm{F}_{\mathrm{MSY}}$, and Figures 31.25 and 31.26 the forecast including FMSY ranges.

## Additional options

A full management options table is provided in Tables 31.14 and 31.15 shows the additional options.

Implementing the previous (now outdated) management plan for this stock with $\mathrm{F}_{\text {MGT }}=0.27$ leads to a total yield (catch) of 1391 t in 2020, and an SSB of 4418 t in 2021. Implementing the MSY approach with $F_{M S Y}=0.29$ leads to a total yield of 1478 t in 2020, and an SSB of 4334 t in 2021.

### 33.5 Biological reference points

The most recent reference points for this stock were developed by WKMSYREF4 in 2015 (ICES, 2016b) and are presented in the table below.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 2900 t | The 5th percentile of the distribution of SSB when fishing at $\mathrm{F}_{\text {MSY }}$ (0.29) with no error. | ICES <br> (2016a) |
|  | $\mathrm{F}_{\mathrm{MSY}}$ | 0.29 | The peak of the median landings yield curve. | ICES <br> (2016a) |
|  | $\mathrm{F}_{\text {MSY }}$ lower | 0.16 | Minimum F which produces at least $95 \%$ of maximum yield. | ICES <br> (2016a) |
|  | $\mathrm{F}_{\text {MSY }}$ upper | 0.34 | Maximum F which produces at least $95 \%$ of maximum yield. | ICES <br> (2016a) |
| Precautionary approach | $\mathrm{Bl}_{\text {lim }}$ | 2000 t | Rounded $\mathrm{B}_{\mathrm{pa}} / 1.4$. | ICES <br> (2016a) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 2900 t | Rounded $\mathrm{B}_{\text {loss }}$ (1999 year class). Lowest SSB with high recruitment. | ICES <br> (2016a) |
|  | $\mathrm{F}_{\text {lim }}$ | 0.44 | Segmented regression simulation of recruitment with $\mathrm{B}_{\mathrm{lim}}$ as the breakpoint and no error. | ICES <br> (2016a) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.32 | $\mathrm{F}_{\text {lim }} \times \exp (-1.645 \times \sigma) ; \sigma=0.2$. | ICES <br> (2016a) |
| Previous management plan | $\mathrm{SSB}_{\text {MGT }}$ | Not defined |  |  |
|  | $\mathrm{F}_{\text {MGT }}$ | 0.27 |  | EU (2007) |

### 33.6 Management plan

The European Commission implemented a management plan for the recovery of the stock early in 2007 (Council Regulation (EC) No 509/2007). The management plan has not been formally evaluated, but the working group concluded that: The long-term management target ( $\mathrm{F}_{\text {мGт }}=$ 0.27 ) is precautionary in the sense that it ensures that there is a less than $5 \%$ chance of SSB declining below previously observed levels, as well as maintaining yield within $10 \%$ of MSY (WGCSE note: long-term yield at $F_{M A X}$ ) (working group, 2005; working group, 2006).

This management plan has not been used in recent years and the ICES advice has been based on the MSY approach, targeting Fmsy.

The management plan (Council Regulation (EC) No 509/2007) is no longer in force since 2019 and has been repealed by an EU multiannual plan for stocks fished in the Western Waters and adjacent waters (Regulation (EU) 2019/472, EU, 2019b) which aims at targeting MSY.

### 33.7 Uncertainties in assessment and forecast

The methodology provided is as robust as possible, and does not currently appear to suffer from a serious retrospective pattern.

## Discarding

Discarding is considered to be negligible in this fishery, averaging only $0.26 \%$ of total international catch weight in 2018. Nevertheless, a time-series of available discards information raised to the fleet level should be developed to deal with potential future discard issues effectively and improve estimates of total mortality. The landings obligation was implemented during 20162018 with a discard plan, and seemed to have reduced the already low discards even more. The landings advice has been topped up with the available discard information to give catch advice so developing a time-series of discard information appears to be less urgent than in the past.

## Surveys

The assessment methodology includes two survey indices. The UK-Q1SWBeam survey added to the assessment in 2012 covers the entire management area, providing fishery-independent tuning information for the entire age range used in the assessment. Therefore, the assessment now relies much less on the commercial tuning information and is less susceptible to localised exploitation by the fishery. However, there is still some uncertainty with respect to the precision of this information. Consequently, commercial tuning information is still used in the assessment to maintain the balance between accuracy and precision required by management. Survey information for the recruiting year class remains temporally variable and is not used in the forecast for this reason.

## Sampling

Age and length sampling for this stock is mostly adequate. Age data from the largest two sectors operating in this fishery (UK and France, together taking 95\% of landings) are included in the assessment and the sampled fleets comprise $93 \%$ of total landings. French age data between 2009 and 2014 were insufficient at older ages to raise the length compositions, and therefore UK age data were used to cover the larger fish.

There are only very limited discard age samples but due to very low discarding this does not impose a problem on the assessment or forecast.

## Consistency

The assessment for this stock was last benchmarked in 2012 and an inter-benchmark was held in 2015. The 2019 assessment is consistent with the previous assessment conducted in recent years. Temporal trends in recruitment, SSB and F estimates were virtually identical.

### 33.8 Recommendation for the next benchmark

There is no requirement to benchmark this stock in the short term.
The XSA assessment uses a taper range of 15 years for the tuning indices, effectively downweighting older tuning data and removing data older than 15 years altogether. As tuning timeseries get longer, potentially important information, in particular from the scientific surveys, might get lost in the process. Therefore, a re-evaluation of assessment parametrisation should be considered.

Lpue estimates for the UK-CBT and UK-COT fleets should be closely monitored to avoid the recurrence of inaccuracies in commercial tuning information observed at the 2014 and 2015 working groups. Minor retrospective patterns in stock status and fishing mortality estimates have begun to re-merge but are expected to stabilise as the duration of the lpue time-series increases in future. The rescaling observed in the 2018 and 2019 assessments can be explained by underlying data. Consequently, the next benchmark should evaluate the temporal stability of the retrospective patterns and determine whether the assessment settings need to be revised.

The UK-COT effort has been in continuous decline and reported no activity in 2016 and subsequently, due to a new data base system cannot be replicated anymore. Consequently, a benchmark could investigate the removal of commercial tuning information altogether from the assessment.

As the time-series on discards is increasing a future benchmark might look into including discard estimates in the assessment and estimating historical discards. As of now, discards are very low and due to the implementation of the landing obligation in 2016 unlikely to become a problem in the future.

### 33.9 Management considerations

France provided discard estimates for the first time at the 2016 working group. Discard estimates from France are higher than from the other countries.

Plaice is taken as bycatch in this fishery, and therefore management advice for sole must also take into account the advice for plaice. Anglerfish, cuttlefish, and lemon sole are also important bycatches in this fishery.

### 33.10 Ecosystem considerations and changes in the environment

See stock annex.

### 33.11 Regulations and their effects

Management of this stock is mainly by TAC. In 2005, effort restrictions were implemented for beam trawlers and entangling gears targeting sole in this fishery to enforce the TAC and improve data quality. The effort restrictions were included in the 2007 management plan (EU, 2007) and
are continued in the EU multiannual plan (EU, 2019b). The effort restrictions limit the numbers of days at sea for vessels in 7 .e using beam trawls ( $\geq 80 \mathrm{~mm}$ mesh size) and static nets $(\leq 120 \mathrm{~mm}$ mesh size). The limits for effort are set annually in the EU council with the TAC and apply only for vessels which catch more than 300 kg of sole annually.

Mesh restrictions for towed gears are set to 80 mm codends, which correspond well with the minimum landing size of sole at 24 cm ( 25 cm for Belgian vessels since December 2017).

### 33.12 References

EC. 2007. Council Regulation (EC) No. 509/2007 of 7 May 2007 establishing a multi-annual plan for the sustainable exploitation of the stock of sole in the Western Channel. Official Journal of the European Union, L 122/7. http://data.europa.eu/eli/reg/2007/509/oj.

EU. 2015. Commission Delegated Regulation (EU) 2015/2438 of 12 October 2015 establishing a discard plan for certain demersal fisheries in north-western waters. OJ L 336, 23.12.2015, p. 29-35. http://eur-lex.europa.eu/eli/reg del/2015/2438/oj.

EU. 2016. Commission Delegated Regulation (EU) 2016/2375 of 12 October 2016 establishing a discard plan for certain demersal fisheries in North-Western waters. C/2016/6439, OJ L 352, 23.12.2016, p. 39-47. http://eur-lex.europa.eu/eli/reg_del/2016/2375/oj.

EU. 2018a. Commission Delegated Regulation (EU) 2018/46 of 20 October 2017 establishing a discard plan for certain demersal and deep sea fisheries in North-Western waters for the year 2018. C/2017/6990, OJ L 7, 12.1.2018, p. 13-20. http://data.europa.eu/eli/reg_del/2018/46/oj.

EU. 2018b. Council Regulation (EU) 2018/120 of 23 January 2018 fixing for 2018 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters, and amending Regulation (EU) 2017/127. OJ L 27, 31.1.2018, p. 1168. http://data.europa.eu/eli/reg/2018/120/oj.

EU. 2019a. Council Regulation (EU) 2019/124 of 30 January 2019 fixing for 2019 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters. ST/15733/2018/INIT. OJ L 29, 31.1.2019, p. 1-166. http://data.europa.eu/eli/reg/2019/124/oj.

EU. 2019b. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. PE/78/2018/REV/1. OJ L 83, 25.3.2019, p. 1-17. http://data.europa.eu/eli/reg/2019/472/oj.

ICES. 2012. Report of the Benchmark Workshop on Flatfish Species and Anglerfish (WKFLAT), 1-8 March 2012, Bilbao, Spain. ICES CM 2012/ACOM:46. 283 pp.

ICES. 2015. Report of the Second Inter-Benchmark Protocol on West of Channel Flatfish (IBPWCFlat2), June-September 2015, By correspondence. ICES CM 2015/ACOM:55. 142 pp.

ICES. 2016a. EU request to ICES to provide Fmsy ranges for selected stocks in ICES subareas 5 to 10. In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.4.1.
ICES. 2016b. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

Table 31.1. Sole in Division 7.e. History of commercial catch and landings. All weights are in tonnes.

| Year | Landings |  |  |  |  |  |  | Discards* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Belgium | France | Netherlands | Ireland | UK and Channel Islands | Unallocated* | Total* |  |
| 1974 |  | 323 |  |  |  | 104 | 427 |  |
| 1975 | 3 | 271 |  |  | 217 |  | 491 |  |
| 1976 | 4 | 352 |  |  | 260 |  | 616 |  |
| 1977 | 3 | 331 |  |  | 272 |  | 606 |  |
| 1978 | 4 | 384 |  |  | 453 | 20 | 861 |  |
| 1979 | 1 | 515 |  |  | 665 |  | 1181 |  |
| 1980 | 45 | 447 |  | 13 | 764 |  | 1269 |  |
| 1981 | 16 | 415 | 1 |  | 788 | -5 | 1215 |  |
| 1982 | 98 | 321 |  |  | 1028 | -1 | 1446 |  |
| 1983 | 47 | 405 | 3 |  | 1043 |  | 1498 |  |
| 1984 | 48 | 421 |  |  | 901 |  | 1370 |  |
| 1985 | 58 | 130 |  |  | 911 | 310 | 1409 |  |
| 1986 | 62 | 467 |  |  | 840 | 50 | 1419 |  |
| 1987 | 48 | 432 |  |  | 632 | 168 | 1280 |  |
| 1988 | 67 | 98 |  |  | 784 | 495 | 1444 |  |
| 1989 | 69 | 112 | 6 |  | 613 | 590 | 1390 |  |
| 1990 | 41 | 81 |  |  | 636 | 556 | 1315 |  |
| 1991 | 35 | 325 |  |  | 477 | 15 | 852 |  |
| 1992 | 41 | 267 |  |  | 468 | 119 | 895 |  |
| 1993 | 59 | 236 |  |  | 498 | 111 | 904 |  |
| 1994 | 33 | 257 |  |  | 546 | -38 | 800 |  |
| 1995 | 21 | 294 |  |  | 565 | -24 | 856 |  |
| 1996 | 8 | 297 |  |  | 428 | 91 | 833 |  |
| 1997 | 13 | 348 |  | 1 | 496 | 91 | 949 |  |
| 1998 | 40 | 343 |  |  | 389 | 108 | 880 |  |
| 1999 | 13 |  |  |  | 396 | 548 | 957 |  |
| 2000 | 4 | 241 |  |  | 413 | 256 | 914 |  |


| Year | Landings |  |  |  |  |  |  | Discards* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Belgium | France | Netherlands | Ireland | UK and Channel Islands | Unallocated* | Total* |  |
| 2001 | 19 | 224 |  |  | 407 | 419 | 1069 |  |
| 2002 | 33 | 198 |  |  | 309 | 566 | 1106 |  |
| 2003 | 1 | 363 |  | 1 | 255 | 458 | 1078 |  |
| 2004 | 7 | 302 |  |  | 185 | 581 | 1075 |  |
| 2005 | 26 | 406 |  |  | 527 | 80 | 1039 |  |
| 2006 | 32 | 357 |  |  | 572 | 61 | 1022 |  |
| 2007 | 34 | 384 |  |  | 536 | 61 | 1015 |  |
| 2008 | 28 | 312 |  | 0 | 472 | 96 | 908 |  |
| 2009 | 17 | 386 |  |  | 381 | -83 | 701 |  |
| 2010 | 17 | 375 |  |  | 370 | -64 | 698 |  |
| 2011 | 22 | 424 |  |  | 431 | -76 | 801 |  |
| 2012 | 39 | 325 |  | 0 | 506 | 2 | 872 | 2 |
| 2013 | 30 | 319 |  |  | 540 | -6 | 883 | 1 |
| 2014 | 25 | 351 |  |  | 509 | -1 | 884 | 10 |
| 2015 | 42 | 245 |  | 0 | 490 | -3 | 774 | 54 |
| 2016 | 46 | 245 |  |  | 623 | -1 | 913 | 10 |
| 2017 | 56 | 198 |  | < 1 | 746 | 7 | 1007 | 4 |
| 2018^ | 68 | 217 | 0 | <1 | 789 | 1 | 1075 | 3 |

*ICES estimate.
${ }^{\wedge}$ Preliminary.

Table 31.2. Sole in Division 7.e. Landings numbers-at-age (thousands).

| YEAR\AGE | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 89 | 322 | 80 | 148 | 210 | 21 | 50 | 26 | 20 | 9 | 63 | 1037 |
| 1970 | 53 | 232 | 322 | 90 | 83 | 112 | 13 | 35 | 52 | 22 | 113 | 1127 |
| 1971 | 51 | 200 | 246 | 198 | 65 | 80 | 156 | 10 | 35 | 54 | 113 | 1207 |
| 1972 | 146 | 412 | 167 | 115 | 112 | 14 | 25 | 134 | 38 | 54 | 106 | 1323 |
| 1973 | 71 | 396 | 433 | 89 | 99 | 120 | 17 | 52 | 30 | 4 | 136 | 1446 |
| 1974 | 45 | 349 | 220 | 178 | 71 | 80 | 43 | 32 | 24 | 55 | 106 | 1202 |
| 1975 | 82 | 567 | 170 | 199 | 115 | 28 | 53 | 26 | 22 | 24 | 171 | 1456 |
| 1976 | 167 | 419 | 472 | 161 | 135 | 92 | 46 | 58 | 51 | 14 | 213 | 1830 |
| 1977 | 426 | 318 | 384 | 206 | 102 | 70 | 74 | 10 | 24 | 32 | 159 | 1804 |
| 1978 | 250 | 1123 | 347 | 214 | 189 | 103 | 72 | 77 | 38 | 27 | 203 | 2644 |
| 1979 | 227 | 803 | 811 | 250 | 229 | 174 | 103 | 90 | 104 | 28 | 290 | 3108 |
| 1980 | 175 | 559 | 497 | 630 | 126 | 183 | 140 | 65 | 56 | 130 | 342 | 2902 |
| 1981 | 245 | 806 | 651 | 467 | 389 | 179 | 126 | 76 | 58 | 55 | 211 | 3262 |
| 1982 | 128 | 1451 | 916 | 553 | 352 | 240 | 136 | 113 | 81 | 61 | 294 | 4324 |
| 1983 | 91 | 753 | 1573 | 583 | 351 | 267 | 294 | 119 | 73 | 37 | 262 | 4401 |
| 1984 | 333 | 663 | 826 | 758 | 325 | 204 | 129 | 152 | 54 | 28 | 255 | 3727 |
| 1985 | 287 | 1700 | 756 | 469 | 585 | 179 | 97 | 103 | 85 | 29 | 125 | 4414 |
| 1986 | 246 | 1618 | 971 | 421 | 321 | 336 | 84 | 75 | 90 | 74 | 127 | 4363 |
| 1987 | 487 | 808 | 1090 | 427 | 204 | 224 | 229 | 47 | 50 | 41 | 162 | 3770 |
| 1988 | 443 | 1438 | 596 | 728 | 374 | 153 | 162 | 109 | 39 | 50 | 171 | 4262 |
| 1989 | 390 | 871 | 1233 | 497 | 509 | 225 | 110 | 107 | 113 | 48 | 214 | 4316 |
| 1990 | 341 | 902 | 581 | 553 | 244 | 264 | 143 | 103 | 75 | 85 | 235 | 3525 |
| 1991 | 450 | 415 | 482 | 289 | 220 | 93 | 111 | 68 | 37 | 31 | 145 | 2341 |
| 1992 | 316 | 1434 | 417 | 297 | 115 | 112 | 61 | 74 | 26 | 23 | 90 | 2964 |
| 1993 | 209 | 704 | 1107 | 350 | 219 | 151 | 78 | 60 | 56 | 31 | 79 | 3045 |
| 1994 | 97 | 657 | 558 | 558 | 112 | 106 | 49 | 57 | 44 | 50 | 99 | 2388 |
| 1995 | 95 | 308 | 629 | 427 | 411 | 131 | 101 | 61 | 33 | 18 | 142 | 2356 |
| 1996 | 365 | 445 | 364 | 298 | 235 | 257 | 68 | 61 | 49 | 37 | 143 | 2321 |


| YEAR\AGE | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 216 | 831 | 724 | 325 | 180 | 194 | 173 | 44 | 20 | 40 | 88 | 2835 |
| 1998 | 265 | 606 | 536 | 336 | 209 | 151 | 80 | 127 | 35 | 34 | 162 | 2543 |
| 1999 | 280 | 915 | 500 | 398 | 255 | 114 | 103 | 54 | 107 | 25 | 123 | 2874 |
| 2000 | 307 | 599 | 751 | 367 | 229 | 107 | 53 | 68 | 51 | 88 | 91 | 2710 |
| 2001 | 145 | 1401 | 531 | 497 | 268 | 178 | 100 | 55 | 43 | 42 | 159 | 3419 |
| 2002 | 332 | 1251 | 843 | 387 | 322 | 129 | 105 | 94 | 33 | 18 | 85 | 3599 |
| 2003 | 598 | 835 | 953 | 645 | 130 | 74 | 50 | 58 | 63 | 14 | 61 | 3482 |
| 2004 | 398 | 1080 | 448 | 445 | 526 | 164 | 116 | 61 | 54 | 35 | 85 | 3412 |
| 2005 | 258 | 468 | 834 | 449 | 366 | 293 | 113 | 80 | 45 | 24 | 96 | 3027 |
| 2006 | 500 | 786 | 472 | 606 | 250 | 224 | 185 | 85 | 56 | 31 | 87 | 3282 |
| 2007 | 201 | 852 | 755 | 293 | 362 | 179 | 130 | 110 | 55 | 27 | 99 | 3062 |
| 2008 | 281 | 752 | 678 | 376 | 163 | 184 | 105 | 71 | 67 | 39 | 89 | 2805 |
| 2009 | 166 | 540 | 385 | 333 | 202 | 66 | 74 | 37 | 50 | 35 | 65 | 1955 |
| 2010 | 68 | 348 | 394 | 329 | 204 | 127 | 49 | 71 | 20 | 34 | 78 | 1723 |
| 2011 | 91 | 499 | 476 | 405 | 233 | 156 | 80 | 39 | 34 | 28 | 93 | 2136 |
| 2012 | 31 | 227 | 525 | 400 | 355 | 231 | 137 | 67 | 44 | 39 | 124 | 2180 |
| 2013 | 120 | 324 | 483 | 595 | 280 | 214 | 147 | 98 | 48 | 23 | 110 | 2441 |
| 2014 | 198 | 320 | 466 | 426 | 410 | 168 | 112 | 79 | 61 | 27 | 97 | 2364 |
| 2015 | 177 | 329 | 395 | 336 | 261 | 206 | 115 | 78 | 45 | 30 | 82 | 2054 |
| 2016 | 92 | 420 | 469 | 276 | 249 | 242 | 189 | 67 | 50 | 33 | 107 | 2194 |
| 2017 | 123 | 1188 | 334 | 307 | 277 | 130 | 94 | 41 | 36 | 129 | 78 | 2737 |
| 2018 | 80 | 446 | 410 | 272 | 339 | 156 | 242 | 99 | 82 | 221 | 154 | 2501 |

Table 31.3. Sole in Division 7.e. Landings weights-at-age (kg).

| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 0.188 | 0.245 | 0.332 | 0.329 | 0.367 | 0.522 | 0.455 | 0.463 | 0.606 | 0.648 | 0.661 |
| 1970 | 0.188 | 0.224 | 0.295 | 0.315 | 0.355 | 0.436 | 0.5 | 0.444 | 0.514 | 0.53 | 0.596 |
| 1971 | 0.151 | 0.222 | 0.296 | 0.367 | 0.35 | 0.359 | 0.431 | 0.455 | 0.476 | 0.388 | 0.654 |
| 1972 | 0.194 | 0.227 | 0.272 | 0.369 | 0.408 | 0.458 | 0.496 | 0.402 | 0.454 | 0.509 | 0.601 |
| 1973 | 0.203 | 0.224 | 0.262 | 0.311 | 0.382 | 0.415 | 0.46 | 0.467 | 0.538 | 0.655 | 0.562 |
| 1974 | 0.183 | 0.224 | 0.281 | 0.379 | 0.434 | 0.372 | 0.465 | 0.476 | 0.488 | 0.475 | 0.732 |
| 1975 | 0.178 | 0.21 | 0.293 | 0.351 | 0.395 | 0.427 | 0.487 | 0.58 | 0.638 | 0.525 | 0.663 |
| 1976 | 0.17 | 0.218 | 0.287 | 0.324 | 0.391 | 0.455 | 0.414 | 0.476 | 0.479 | 0.585 | 0.629 |
| 1977 | 0.197 | 0.249 | 0.303 | 0.357 | 0.4 | 0.503 | 0.464 | 0.518 | 0.485 | 0.553 | 0.683 |
| 1978 | 0.178 | 0.239 | 0.3 | 0.387 | 0.435 | 0.374 | 0.482 | 0.485 | 0.484 | 0.535 | 0.665 |
| 1979 | 0.189 | 0.239 | 0.33 | 0.427 | 0.464 | 0.472 | 0.481 | 0.57 | 0.527 | 0.574 | 0.732 |
| 1980 | 0.189 | 0.254 | 0.343 | 0.389 | 0.525 | 0.56 | 0.609 | 0.646 | 0.655 | 0.6 | 0.783 |
| 1981 | 0.174 | 0.225 | 0.321 | 0.381 | 0.477 | 0.514 | 0.533 | 0.598 | 0.619 | 0.708 | 0.66 |
| 1982 | 0.214 | 0.209 | 0.278 | 0.347 | 0.426 | 0.498 | 0.51 | 0.523 | 0.526 | 0.564 | 0.663 |
| 1983 | 0.187 | 0.25 | 0.271 | 0.306 | 0.388 | 0.417 | 0.473 | 0.53 | 0.608 | 0.551 | 0.665 |
| 1984 | 0.21 | 0.243 | 0.306 | 0.381 | 0.391 | 0.481 | 0.542 | 0.562 | 0.604 | 0.726 | 0.643 |
| 1985 | 0.163 | 0.226 | 0.298 | 0.36 | 0.391 | 0.472 | 0.523 | 0.534 | 0.522 | 0.588 | 0.822 |
| 1986 | 0.174 | 0.237 | 0.297 | 0.354 | 0.407 | 0.456 | 0.502 | 0.544 | 0.583 | 0.618 | 0.703 |
| 1987 | 0.174 | 0.245 | 0.31 | 0.37 | 0.425 | 0.474 | 0.518 | 0.557 | 0.59 | 0.618 | 0.665 |
| 1988 | 0.17 | 0.244 | 0.312 | 0.375 | 0.432 | 0.484 | 0.531 | 0.572 | 0.608 | 0.639 | 0.694 |
| 1989 | 0.167 | 0.222 | 0.275 | 0.326 | 0.375 | 0.422 | 0.467 | 0.51 | 0.551 | 0.59 | 0.692 |
| 1990 | 0.217 | 0.272 | 0.324 | 0.372 | 0.419 | 0.461 | 0.501 | 0.538 | 0.571 | 0.601 | 0.669 |
| 1991 | 0.182 | 0.255 | 0.323 | 0.386 | 0.445 | 0.499 | 0.549 | 0.594 | 0.634 | 0.669 | 0.741 |
| 1992 | 0.166 | 0.238 | 0.305 | 0.366 | 0.423 | 0.474 | 0.52 | 0.561 | 0.597 | 0.627 | 0.683 |
| 1993 | 0.146 | 0.209 | 0.268 | 0.324 | 0.376 | 0.425 | 0.47 | 0.513 | 0.551 | 0.587 | 0.672 |
| 1994 | 0.183 | 0.241 | 0.295 | 0.347 | 0.396 | 0.442 | 0.484 | 0.524 | 0.561 | 0.595 | 0.671 |
| 1995 | 0.192 | 0.248 | 0.301 | 0.351 | 0.397 | 0.441 | 0.481 | 0.518 | 0.552 | 0.583 | 0.652 |
| 1996 | 0.214 | 0.262 | 0.308 | 0.354 | 0.399 | 0.442 | 0.484 | 0.524 | 0.564 | 0.602 | 0.694 |


| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 0.186 | 0.244 | 0.3 | 0.354 | 0.406 | 0.455 | 0.503 | 0.548 | 0.592 | 0.633 | 0.734 |
| 1998 | 0.191 | 0.247 | 0.3 | 0.35 | 0.397 | 0.441 | 0.482 | 0.52 | 0.555 | 0.586 | 0.661 |
| 1999 | 0.208 | 0.257 | 0.303 | 0.347 | 0.389 | 0.429 | 0.468 | 0.503 | 0.536 | 0.567 | 0.637 |
| 2000 | 0.202 | 0.258 | 0.31 | 0.358 | 0.401 | 0.441 | 0.476 | 0.508 | 0.535 | 0.558 | 0.647 |
| 2001 | 0.203 | 0.245 | 0.287 | 0.326 | 0.365 | 0.402 | 0.438 | 0.472 | 0.505 | 0.537 | 0.616 |
| 2002 | 0.181 | 0.236 | 0.29 | 0.342 | 0.391 | 0.439 | 0.485 | 0.529 | 0.57 | 0.61 | 0.706 |
| 2003 | 0.173 | 0.241 | 0.306 | 0.367 | 0.425 | 0.479 | 0.53 | 0.577 | 0.62 | 0.66 | 0.746 |
| 2004 | 0.176 | 0.23 | 0.282 | 0.334 | 0.385 | 0.435 | 0.485 | 0.534 | 0.582 | 0.629 | 0.757 |
| 2005 | 0.18 | 0.236 | 0.29 | 0.343 | 0.394 | 0.444 | 0.493 | 0.54 | 0.586 | 0.63 | 0.747 |
| 2006 | 0.169 | 0.228 | 0.282 | 0.333 | 0.381 | 0.424 | 0.464 | 0.501 | 0.533 | 0.562 | 0.672 |
| 2007 | 0.183 | 0.244 | 0.299 | 0.35 | 0.395 | 0.436 | 0.471 | 0.501 | 0.526 | 0.546 | 0.616 |
| 2008 | 0.197 | 0.245 | 0.292 | 0.337 | 0.382 | 0.425 | 0.468 | 0.509 | 0.549 | 0.588 | 0.652 |
| 2009 | 0.176 | 0.252 | 0.322 | 0.385 | 0.443 | 0.494 | 0.54 | 0.579 | 0.612 | 0.639 | 0.703 |
| 2010 | 0.169 | 0.258 | 0.339 | 0.412 | 0.476 | 0.532 | 0.58 | 0.619 | 0.65 | 0.673 | 0.699 |
| 2011 | 0.2 | 0.261 | 0.319 | 0.375 | 0.428 | 0.48 | 0.528 | 0.575 | 0.618 | 0.66 | 0.749 |
| 2012 | 0.162 | 0.24 | 0.311 | 0.373 | 0.428 | 0.476 | 0.516 | 0.548 | 0.572 | 0.589 | 0.664 |
| 2013 | 0.172 | 0.228 | 0.283 | 0.337 | 0.389 | 0.439 | 0.489 | 0.536 | 0.583 | 0.628 | 0.74 |
| 2014 | 0.191 | 0.254 | 0.313 | 0.366 | 0.415 | 0.459 | 0.499 | 0.533 | 0.563 | 0.588 | 0.709 |
| 2015 | 0.182 | 0.25 | 0.313 | 0.37 | 0.423 | 0.471 | 0.513 | 0.551 | 0.583 | 0.611 | 0.697 |
| 2016 | 0.215 | 0.282 | 0.345 | 0.401 | 0.453 | 0.499 | 0.541 | 0.576 | 0.606 | 0.631 | 0.72 |
| 2017 | 0.225 | 0.279 | 0.331 | 0.382 | 0.432 | 0.479 | 0.525 | 0.568 | 0.61 | 0.651 | 0.763 |
| 2018 | 0.205 | 0.264 | 0.321 | 0.374 | 0.425 | 0.473 | 0.518 | 0.56 | 0.6 | 0.636 | 0.768 |

Table 31.4. Sole in Division 7.e. Stock weights-at-age (kg).

| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 0.125 | 0.2 | 0.27 | 0.33 | 0.38 | 0.425 | 0.46 | 0.49 | 0.52 | 0.55 | 0.609 |
| 1970 | 0.12 | 0.195 | 0.255 | 0.305 | 0.355 | 0.395 | 0.43 | 0.465 | 0.49 | 0.51 | 0.541 |
| 1971 | 0.09 | 0.17 | 0.24 | 0.295 | 0.345 | 0.39 | 0.42 | 0.445 | 0.47 | 0.49 | 0.544 |
| 1972 | 0.13 | 0.2 | 0.265 | 0.325 | 0.38 | 0.42 | 0.46 | 0.49 | 0.52 | 0.54 | 0.558 |
| 1973 | 0.105 | 0.17 | 0.235 | 0.29 | 0.34 | 0.39 | 0.435 | 0.475 | 0.51 | 0.54 | 0.585 |
| 1974 | 0.125 | 0.2 | 0.265 | 0.32 | 0.37 | 0.41 | 0.455 | 0.49 | 0.515 | 0.53 | 0.571 |
| 1975 | 0.144 | 0.221 | 0.267 | 0.327 | 0.385 | 0.435 | 0.479 | 0.516 | 0.545 | 0.569 | 0.628 |
| 1976 | 0.146 | 0.198 | 0.247 | 0.294 | 0.338 | 0.38 | 0.417 | 0.456 | 0.491 | 0.523 | 0.595 |
| 1977 | 0.156 | 0.221 | 0.278 | 0.332 | 0.382 | 0.425 | 0.462 | 0.497 | 0.527 | 0.553 | 0.629 |
| 1978 | 0.156 | 0.217 | 0.276 | 0.33 | 0.38 | 0.425 | 0.463 | 0.498 | 0.526 | 0.555 | 0.63 |
| 1979 | 0.141 | 0.216 | 0.287 | 0.352 | 0.414 | 0.463 | 0.502 | 0.539 | 0.574 | 0.608 | 0.719 |
| 1980 | 0.125 | 0.206 | 0.288 | 0.36 | 0.436 | 0.513 | 0.575 | 0.62 | 0.65 | 0.674 | 0.714 |
| 1981 | 0.119 | 0.197 | 0.276 | 0.358 | 0.427 | 0.49 | 0.543 | 0.582 | 0.616 | 0.645 | 0.699 |
| 1982 | 0.117 | 0.195 | 0.265 | 0.335 | 0.398 | 0.455 | 0.506 | 0.536 | 0.562 | 0.585 | 0.632 |
| 1983 | 0.12 | 0.195 | 0.25 | 0.307 | 0.365 | 0.42 | 0.475 | 0.52 | 0.57 | 0.615 | 0.709 |
| 1984 | 0.108 | 0.192 | 0.268 | 0.339 | 0.4 | 0.453 | 0.501 | 0.545 | 0.577 | 0.607 | 0.696 |
| 1985 | 0.15 | 0.204 | 0.258 | 0.311 | 0.364 | 0.416 | 0.468 | 0.52 | 0.571 | 0.621 | 0.79 |
| 1986 | 0.14 | 0.206 | 0.268 | 0.326 | 0.381 | 0.432 | 0.48 | 0.524 | 0.564 | 0.601 | 0.691 |
| 1987 | 0.137 | 0.21 | 0.278 | 0.341 | 0.398 | 0.45 | 0.497 | 0.538 | 0.574 | 0.605 | 0.659 |
| 1988 | 0.131 | 0.208 | 0.278 | 0.344 | 0.404 | 0.459 | 0.508 | 0.552 | 0.591 | 0.624 | 0.687 |
| 1989 | 0.139 | 0.195 | 0.249 | 0.3 | 0.35 | 0.398 | 0.444 | 0.488 | 0.531 | 0.571 | 0.675 |
| 1990 | 0.187 | 0.243 | 0.296 | 0.346 | 0.393 | 0.437 | 0.478 | 0.516 | 0.551 | 0.583 | 0.654 |
| 1991 | 0.144 | 0.219 | 0.29 | 0.355 | 0.416 | 0.473 | 0.524 | 0.572 | 0.614 | 0.652 | 0.731 |
| 1992 | 0.128 | 0.202 | 0.272 | 0.336 | 0.395 | 0.449 | 0.498 | 0.542 | 0.58 | 0.613 | 0.677 |
| 1993 | 0.114 | 0.178 | 0.239 | 0.296 | 0.35 | 0.401 | 0.448 | 0.492 | 0.532 | 0.57 | 0.659 |
| 1994 | 0.153 | 0.212 | 0.268 | 0.322 | 0.372 | 0.419 | 0.463 | 0.505 | 0.543 | 0.578 | 0.659 |
| 1995 | 0.163 | 0.221 | 0.275 | 0.326 | 0.374 | 0.419 | 0.461 | 0.5 | 0.536 | 0.568 | 0.641 |
| 1996 | 0.189 | 0.238 | 0.285 | 0.331 | 0.376 | 0.42 | 0.463 | 0.504 | 0.544 | 0.583 | 0.677 |


| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 0.156 | 0.215 | 0.272 | 0.327 | 0.38 | 0.431 | 0.48 | 0.526 | 0.57 | 0.612 | 0.717 |
| 1998 | 0.162 | 0.22 | 0.274 | 0.325 | 0.374 | 0.419 | 0.462 | 0.501 | 0.537 | 0.571 | 0.65 |
| 1999 | 0.183 | 0.233 | 0.28 | 0.326 | 0.369 | 0.41 | 0.448 | 0.485 | 0.519 | 0.551 | 0.624 |
| 2000 | 0.172 | 0.23 | 0.284 | 0.333 | 0.379 | 0.421 | 0.458 | 0.492 | 0.521 | 0.546 | 0.643 |
| 2001 | 0.181 | 0.224 | 0.266 | 0.307 | 0.346 | 0.384 | 0.42 | 0.455 | 0.489 | 0.521 | 0.602 |
| 2002 | 0.152 | 0.209 | 0.263 | 0.316 | 0.367 | 0.415 | 0.462 | 0.507 | 0.55 | 0.591 | 0.688 |
| 2003 | 0.137 | 0.207 | 0.274 | 0.337 | 0.396 | 0.452 | 0.505 | 0.554 | 0.599 | 0.641 | 0.732 |
| 2004 | 0.149 | 0.203 | 0.256 | 0.308 | 0.36 | 0.41 | 0.46 | 0.509 | 0.557 | 0.605 | 0.734 |
| 2005 | 0.152 | 0.208 | 0.263 | 0.316 | 0.368 | 0.419 | 0.468 | 0.516 | 0.562 | 0.607 | 0.726 |
| 2006 | 0.138 | 0.197 | 0.254 | 0.306 | 0.355 | 0.4 | 0.442 | 0.479 | 0.514 | 0.544 | 0.661 |
| 2007 | 0.151 | 0.214 | 0.272 | 0.325 | 0.373 | 0.416 | 0.454 | 0.486 | 0.514 | 0.536 | 0.614 |
| 2008 | 0.172 | 0.221 | 0.268 | 0.315 | 0.36 | 0.404 | 0.447 | 0.489 | 0.529 | 0.569 | 0.64 |
| 2009 | 0.136 | 0.215 | 0.287 | 0.354 | 0.415 | 0.469 | 0.518 | 0.56 | 0.596 | 0.626 | 0.698 |
| 2010 | 0.121 | 0.215 | 0.3 | 0.376 | 0.445 | 0.505 | 0.557 | 0.6 | 0.636 | 0.663 | 0.696 |
| 2011 | 0.169 | 0.231 | 0.29 | 0.347 | 0.402 | 0.454 | 0.504 | 0.552 | 0.597 | 0.639 | 0.738 |
| 2012 | 0.12 | 0.202 | 0.276 | 0.343 | 0.402 | 0.453 | 0.497 | 0.532 | 0.561 | 0.581 | 0.664 |
| 2013 | 0.144 | 0.2 | 0.256 | 0.31 | 0.363 | 0.414 | 0.464 | 0.513 | 0.56 | 0.606 | 0.729 |
| 2014 | 0.157 | 0.223 | 0.284 | 0.34 | 0.391 | 0.438 | 0.48 | 0.517 | 0.549 | 0.576 | 0.706 |
| 2015 | 0.147 | 0.217 | 0.282 | 0.342 | 0.397 | 0.448 | 0.493 | 0.533 | 0.568 | 0.598 | 0.692 |
| 2016 | 0.178 | 0.248 | 0.313 | 0.373 | 0.427 | 0.476 | 0.519 | 0.557 | 0.59 | 0.617 | 0.714 |
| 2017 | 0.197 | 0.252 | 0.305 | 0.357 | 0.407 | 0.455 | 0.501 | 0.546 | 0.588 | 0.630 | 0.749 |
| 2018 | 0.174 | 0.235 | 0.293 | 0.348 | 0.400 | 0.450 | 0.496 | 0.540 | 0.580 | 0.618 | 0.760 |

Table 31.5. Sole in Division 7.e. Landings, effort and mean standardised lpue for the UK commercial fleets.

| Fleet | Year | Effort [days] | Landings [tonnes] | Lpue [tonnes/1000 days] | means standardised Ipue |
| :---: | :---: | :---: | :---: | :---: | :---: |
| UK-CBT<24 m | 1988 | 2527 | 293 | 115.97 | 1.96 |
|  | 1989 | 1956 | 162 | 83.06 | 1.4 |
|  | 1990 | 1958 | 179 | 91.51 | 1.54 |
|  | 1991 | 1458 | 134 | 92.22 | 1.56 |
|  | 1992 | 1342 | 142 | 106.22 | 1.79 |
|  | 1993 | 1432 | 154 | 107.71 | 1.82 |
|  | 1994 | 2241 | 161 | 71.97 | 1.21 |
|  | 1995 | 2017 | 134 | 66.28 | 1.12 |
|  | 1996 | 1999 | 106 | 52.99 | 0.89 |
|  | 1997 | 1991 | 132 | 66.3 | 1.12 |
|  | 1998 | 2357 | 99 | 42.12 | 0.71 |
|  | 1999 | 2518 | 115 | 45.7 | 0.77 |
|  | 2000 | 2913 | 134 | 45.85 | 0.77 |
|  | 2001 | 3746 | 148 | 39.57 | 0.67 |
|  | 2002 | 3482 | 110 | 31.55 | 0.53 |
|  | 2003 | 3785 | 93 | 24.44 | 0.41 |
|  | 2004 | 3512 | 64 | 18.12 | 0.31 |
|  | 2005 | 3305 | 191 | 57.72 | 0.97 |
|  | 2006 | 3277 | 224 | 68.27 | 1.15 |
|  | 2007 | 4027 | 225 | 55.77 | 0.94 |
|  | 2008 | 4629 | 213 | 45.94 | 0.78 |
|  | 2009 | 4040 | $185$ | 45.85 | 0.77 |
|  | 2010 | 4727 | 201 | 42.42 | 0.72 |
|  | 2011 | 5913 | 258 | 43.65 | 0.74 |
|  | 2012 | 7188 | 314 | 43.65 | 0.74 |
|  | 2013 | 6322 | 329 | 52.02 | 0.88 |
|  | 2014 | 5870 | 308 | 52.54 | 0.89 |
|  | 2015 | 6260 | 310 | 49.54 | 0.84 |
|  | 2016 | 6114 | 355 | 58.1 | 0.98 |
|  | 2017 | 6556 | 400 | 61.07 | 1.03 |
|  | 2018 | 6366 | $386$ | 60.66 | 1.02 |
| UK-CBT>24 m | 1988 | 2971 | 391 | 131.77 | 2.92 |
|  | 1989 | 3938 | 340 | 86.37 | 1.91 |
|  | 1990 | 3518 | 314 | 89.12 | 1.97 |
|  | 1991 | 2412 | 206 | 85.47 | 1.89 |
|  | 1992 | 1993 | 197 | 98.63 | 2.18 |
|  | 1993 | 2678 | 194 | 72.54 | 1.61 |


| Fleet | Year | Effort [days] | Landings [tonnes] | Lpue [tonnes/1000 days] | means standardised Ipue |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1994 | 4574 | 236 | 51.5 | 1.14 |
|  | 1995 | 4917 | 257 | 52.3 | 1.16 |
|  | 1996 | 5592 | 178 | 31.84 | 0.7 |
|  | 1997 | 5377 | 199 | 37.1 | 0.82 |
|  | 1998 | 4945 | 164 | 33.19 | 0.73 |
|  | 1999 | 4512 | 141 | 31.32 | 0.69 |
|  | 2000 | 5237 | 151 | 28.84 | 0.64 |
|  | 2001 | 5874 | 142 | 24.11 | 0.53 |
|  | 2002 | 5957 | 104 | 17.51 | 0.39 |
|  | 2003 | 6810 | 94 | 13.78 | 0.31 |
|  | 2004 | 7100 | 69 | 9.66 | 0.21 |
|  | 2005 | 6684 | 236 | 35.27 | 0.78 |
|  | 2006 | 6595 | 236 | 35.79 | 0.79 |
|  | 2007 | 5594 | 196 | 35.1 | 0.78 |
|  | 2008 | 4924 | 154 | 31.36 | 0.69 |
|  | 2009 | 3523 | 115 | 32.66 | 0.72 |
|  | 2010 | 3064 | 94 | 30.64 | 0.68 |
|  | 2011 | 2790 | 92 | 32.95 | 0.73 |
|  | 2012 | 2609 | 86 | 33.01 | 0.73 |
|  | 2013 | 2444 | 93 | 38.13 | 0.84 |
|  | 2014 | 2900 | 104 | 35.95 | 0.8 |
|  | 2015 | 3039 | 101 | 33.12 | 0.73 |
|  | 2016 | 4064 | 166 | 40.79 | 0.9 |
|  | 2017 | 4556 | 207 | 45.41 | 1.01 |
|  | 2018 | 4116 | 231 | 56.17 | 1.23 |
| UK-CBT | 1988 | 5497 | 684 | 124.51 | 2.43 |
|  | 1989 | 5894 | 503 | 85.27 | 1.66 |
|  | 1990 | 5476 | 493 | 89.97 | 1.76 |
|  | 1991 | 3870 | 341 | 88.02 | 1.72 |
|  | 1992 | 3334 | 339 | 101.69 | 1.99 |
|  | 1993 | 4111 | 349 | 84.79 | 1.66 |
|  | 1994 | 6814 | 397 | 58.23 | 1.14 |
|  | 1995 | 6935 | 391 | 56.37 | 1.1 |
|  | 1996 | 7591 | 284 | 37.41 | 0.73 |
|  | 1997 | 7368 | 331 | 44.99 | 0.88 |
|  | 1998 | 7302 | 263 | 36.07 | 0.7 |
|  | 1999 | 7031 | 256 | 36.47 | 0.71 |
|  | 2000 | 8150 | 285 | 34.92 | 0.68 |


| Fleet | Year | Effort [days] | Landings [tonnes] | Lpue [tonnes/1000 days] | means standardised Ipue |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2001 | 9620 | 290 | 30.13 | 0.59 |
|  | 2002 | 9439 | 214 | 22.69 | 0.44 |
|  | 2003 | 10596 | 186 | 17.59 | 0.34 |
|  | 2004 | 10612 | 132 | 12.46 | 0.24 |
|  | 2005 | 9990 | 427 | 42.7 | 0.83 |
|  | 2006 | 9873 | 460 | 46.57 | 0.91 |
|  | 2007 | 9621 | 421 | 43.75 | 0.85 |
|  | 2008 | 9552 | 367 | 38.42 | 0.75 |
|  | 2009 | 7563 | 300 | 39.7 | 0.78 |
|  | 2010 | 7791 | 294 | 37.79 | 0.74 |
|  | 2011 | 8703 | 350 | 40.22 | 0.79 |
|  | 2012 | 9797 | 400 | 40.82 | 0.8 |
|  | 2013 | 8767 | 422 | 48.15 | 0.94 |
|  | 2014 | 8769 | 413 | 47.05 | 0.92 |
|  | 2015 | 9298 | 411 | 44.17 | 0.86 |
|  | 2016 | 10178 | 521 | 51.19 | 1 |
|  | 2017 | 11112 | 607 | 54.65 | 1.07 |
|  | 2018 | 10482 | 617 | 58.9 | 1.14 |
| UK-COT | 1988 | 4265 | 29 | 6.77 | 1.43 |
|  | 1989 | 4607 | 28 | 6.18 | 1.31 |
|  | 1990 | 4423 | 26 | 5.97 | 1.27 |
|  | 1991 | 4004 | 14 | 3.39 | 0.72 |
|  | 1992 | 4108 | 12 | 3.02 | 0.64 |
|  | 1993 | 3761 | 15 | 3.95 | 0.84 |
|  | 1994 | 3423 | 18 | 5.27 | 1.12 |
|  | 1995 | 3294 | 13 | 3.99 | 0.84 |
|  | 1996 | 2589 | 12 | 4.83 | 1.02 |
|  | 1997 | 3011 | 15 | 4.96 | 1.05 |
|  | 1998 | 2699 | 11 | 4.22 | 0.89 |
|  | 1999 | 2486 | 13 | 5.16 | 1.09 |
|  | 2000 | 2681 | 11 | 4.11 | 0.87 |
|  | 2001 | 2732 | 13 | 4.9 | 1.04 |
|  | 2002 | 2448 | 9 | 3.66 | 0.78 |
|  | 2003 | 2273 | 8 | 3.31 | 0.7 |
|  | 2004 | 2334 | 6 | 2.46 | 0.52 |
|  | 2005 | 1762 | 12 | 6.86 | 1.45 |
|  | 2006 | 1699 | 8 | 4.57 | 0.97 |
|  | 2007 | 1917 | 9 | 4.9 | 1.04 |


| Fleet | Year | Effort [days] | Landings [tonnes] | Lpue [tonnes/1000 days] | means standardised Ipue |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2008 | 1750 | 7 | 4.26 | 0.9 |
|  | 2009 | 1847 | 10 | 5.36 | 1.14 |
|  | 2010 | 2213 | 10 | 4.53 | 0.96 |
|  | 2011 | 1930 | 8 | 4.08 | 0.86 |
|  | 2012 | 2068 | 12 | 5.96 | 1.26 |
|  | 2013 | 1587 | 8 | 4.96 | 1.05 |
|  | 2014 | 1440 | 8 | 5.56 | 1.18 |
|  | 2015 | 978 | 5 | 4.98 | 1.06 |
|  | 2016 | 0 | 0 | NA | NA |

Note that the lpue time-series for the UK commercial beam-trawl fleet was revised at IBPWCFlat2 due to modifications in the UK e-logbook effort recording system in 2012.

Table 31.6. Sole in Division 7.e. Tuning data file. Not all tuning time-series, years and ages shown here were used in the assessment.

| 104 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK-CBT-late |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 20032018 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1101 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 314 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 10.66126 | 130.7 | 168.87 | 129.96 | 21.43 | 18.32 | 10.28 | 13.49 | 6.67 | 2.19 | 2.06 | 3.35 | 2.82 |  |
| 10.61079 | 146.5 | 61.53 | 53.46 | 75.23 | 11.35 | 14.96 | 7.49 5 | 5.98 | 4.27 | 2.121 | $1.18 \quad 1.8$ | . 89 |  |
| 9.99213 | 210.39 | 326.3 | 132.94 | 155.21 | 132.09 | 27.41 | 32.6 | 22.54 | 14.24 | 8.3 | 5.95 | 4.84 |  |
| 9.89192 | 376.87 | 186.46 | 243.45 | 85.59 | 108.34 | 106.98 | 37.22 | 20.67 | 13.69 | 13.61 | - 6.68 | 2.99 |  |
| 9.61475 | 456.04 | 261.42 | 105.82 | 103.55 | 54.21 | 62.07 | 51.47 | 15.34 | 11.12 | 10.41 | 8.44 | 8.17 |  |
| 9.55107 | 294.03 | 286.06 | 126.1 | 67.89 | 65.42 | 42.34 | 39.54 | 36.27 | 14.54 | 11.8 | 4.3 | 6 |  |
| 7.56283 | 190.03 | 182.63 | 152.83 | 89.59 | 26.02 | 27.9 | 13.23 | 16.1 | 12.91 | 4.85 | 3.74 | 1.92 |  |
| 7.78378 | 80.09 | 179.7 | 157.57 | 101.24 | 51.98 | 25.24 | 22.59 | 8.23 | 16.75 | 25.39 | 7.42 | 3.88 |  |
| 8.70071 | 243.76 | 148.58 | 186.66 | 121.43 | 81.66 | 35.56 | 15.79 | 20.25 | 10.83 | 14.11 | 8.26 | 2.1 |  |
| 9.78759 | 129.79 | 307.88 | 139.02 | 143.59 | 91.49 | 66.22 | 30.49 | 17.81 | 14.83 | 8.55 | 12.25 | 11.03 |  |
| 8.75236 | 81.92 | 242.49 | 288.92 | 134.34 | 93.18 | 72.27 | 44.15 | 24.5 | 10.73 | 9.84 | 8.14 | 9.84 |  |
| 8.7411 | 111.72 | 201.15 | 169.62 | 201.19 | 99.91 | 67.46 | 43.84 | 30.63 | 15.94 | 7.71 | 9.34 | 4.9 |  |
| 9.27543 | 137.05 | 178.21 | 198.83 | 135.74 | 117.19 | 65.74 | 45.95 | 31.78 | - 20.59 | 11.01 | - 5.52 | 5.96 |  |
| 10.17804 | 263.46 | 217.34 | 158.93 | 161.88 | 118.88 | 102.14 | 49.07 | - 45.22 | $22 \quad 21.3$ | - 23.14 | $4 \quad 13.03$ | 35.69 |  |
| 11.13423 | 454.27 | 353.27 | 177.37 | 142.06 | 120.28 | 81.72 | 72.95 | 42.23 | $3 \quad 28.03$ | -16.59 | $9 \quad 11.97$ | $7 \quad 9.63$ |  |
| 10.48248 | 217.63 | 454.82 | 260.75 | 116.59 | 118.4 | 76.79 | 51.54 | 49.36 | - 33.91 | 24.42 | 221.84 | 10.92 |  |
| UK-COT |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 19882016 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1101 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 311 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 4264.71 |  | 30.97 |  | 15.73 |  | 19.29 |  |  | 8.63 |  | 2.55 |  | 2.55 |
|  |  | 1.83 |  | 0.35 |  | 0.76 |  |  |  |  |  |  |  |
| 4607.04 |  | 15.09 |  | 18.34 |  | 9.22 |  |  | 11.75 |  | 4.72 |  | 2.42 |
|  |  | 2.36 |  | 2.01 |  | 1.4 |  |  |  |  |  |  |  |
| 4422.52 |  | 18.3 |  | 12.56 |  | 9.21 |  |  | 6.09 |  | 5.53 |  | 2.08 |
|  |  | 1.83 |  | 1.12 |  | 0.9 |  |  |  |  |  |  |  |
| 4004.37 |  | 10.04 |  | 7.03 |  | 4.12 |  |  | 2.46 |  | 0.96 |  | 1.44 |
|  |  | 0.42 |  | 0.41 |  | 0.23 |  |  |  |  |  |  |  |
| 4107.71 |  | 26.24 |  | 6 |  | 3.6 |  |  | 1.19 |  | 1.14 |  | 0.48 |
|  |  | 0.65 |  | 0.17 |  | 0.09 |  |  |  |  |  |  |  |
| 3761 |  | 12.45 |  | 17.56 |  | 5.38 |  |  | 3.44 |  | 2.49 |  | 1.26 |
|  |  | 1 |  | 0.92 |  | 0.56 |  |  |  |  |  |  |  |
| 3423.03 |  | 12.42 |  | 11.46 |  | 12.35 |  |  | 2.5 |  | 2.6 |  | 1.23 |
|  |  | 1.35 |  | 1.03 |  | 1.18 |  |  |  |  |  |  |  |
| 3294.06 |  | 5.25 |  | 9.75 |  | 6.34 |  |  | 6.17 |  | 1.89 |  | 1.49 |
|  |  | 0.91 |  | 0.52 |  | 0.25 |  |  |  |  |  |  |  |
| 2589.38 |  | 9.47 |  | 6.54 |  | 4.37 |  |  | 3.15 |  | 3.54 |  | 0.95 |
|  |  | 0.76 |  | 0.68 |  | 0.45 |  |  |  |  |  |  |  |
| 3010.66 |  | 15.16 |  | 8.81 |  | 4.78 |  |  | 2.83 |  | 2.9 |  | 2.53 |
|  |  | 0.63 |  | 0.28 |  | 0.43 |  |  |  |  |  |  |  |
| 2698.6 |  | 8.74 |  | 7.58 |  | 4.25 |  |  | 2.49 |  | 1.53 |  | 0.93 |
|  |  | 1.47 |  | 0.31 |  | 0.44 |  |  |  |  |  |  |  |
| 2486.17 |  | 11.56 |  | 5.84 |  | 4.91 |  |  | 2.89 |  | 1.45 |  | 1.46 |
|  |  | 0.74 |  | 1.49 |  | 0.39 |  |  |  |  |  |  |  |


| 2680.63 | 6.67 | 8.41 | 4.03 | 2.64 | 1.24 | 0.59 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.81 | 0.62 | 0.99 |  |  |  |
| 2731.54 | 18.02 | 5.27 | 4.96 | 2.69 | 2.01 | 1.12 |
|  | 0.7 | 0.51 | 0.5 |  |  |  |
| 2448.37 | 9.88 | 6.12 | 2.39 | 2.67 | 1.27 | 0.82 |
|  | 0.33 | 0.2 | 0.25 |  |  |  |
| 2272.9 | 4.61 | 5.87 | 4.8 | 1.04 | 0.85 | 0.49 |
|  | 0.54 | 0.27 | 0.13 |  |  |  |
| 2334.16 | 6.05 | 2.58 | 2.23 | 3.25 | 0.46 | 0.57 |
|  | 0.3 | 0.24 | 0.18 |  |  |  |
| 1762.36 | 6.44 | 9.56 | 3.53 | 4.13 | 3.44 | 0.74 |
|  | 0.9 | 0.58 | 0.45 |  |  |  |
| 1699.49 | 6.93 | 3.27 | 4.13 | 1.36 | 1.63 | 1.75 |
|  | 0.6 | 0.31 | 0.2 |  |  |  |
| 1916.84 | 9.32 | 5.44 | 2.3 | 2.32 | 1.19 | 1.41 |
|  | 1.13 | 0.36 | 0.21 |  |  |  |
| 1750.36 | 5.61 | 4.85 | 2.08 | 1.15 | 1.18 | 0.75 |
|  | 0.75 | 0.7 | 0.32 |  |  |  |
| 1847.2 | 7.97 | 5.47 | 3.92 | 2.17 | 0.64 | 0.83 |
|  | 0.39 | 0.52 | 0.45 |  |  |  |
| 2212.85 | 2.71 | 5.85 | 4.74 | 3.15 | 1.63 | 0.81 |
|  | 0.74 | 0.3 | 0.6 |  |  |  |
| 1930.5 | 6.51 | 3.32 | 3.89 | 2.46 | 1.64 | 0.58 |
|  | 0.31 | 0.37 | 0.19 |  |  |  |
| 2068.16 | 4.24 | 9.16 | 3.97 | 4.06 | 2.3 | 1.76 |
|  | 0.82 | 0.49 | 0.46 |  |  |  |
| 1586.58 | 2.01 | 4.55 | 5.64 | 2.66 | 1.74 | 1.49 |
|  | 0.89 | 0.56 | 0.26 |  |  |  |
| 1440.22 | 2.13 | 3.57 | 2.99 | 3.56 | 1.8 | 1.29 |
|  | 0.9 | 0.68 | 0.34 |  |  |  |
| 977.63 | 1.62 | 1.98 | 1.86 | 1.59 | 1.35 | 0.7 |
|  | 0.5 | 0.42 | 0.25 |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 |  |  |  |

Q1SWBeam-nonoffset
20062018
110.10 .25

127

| 1 | 0 | 13.9827 | 17.7418 | 9.8877 | 19.4529 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 11.9525 | 9.8066 | 10.4549 | 4.74613 | 3.23665 |  |
| 7.00007 | 0.86644 | 1.50309 | 0.42807 | 0.86046 |  |  |
|  | 3.98914 | 0.83462 | 0.39699 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.21454 | 12.3291 | 36.7717 | 16.2021 | 2.0082 |  |
|  | 7.3474 | 2.5642 | 2.7218 | 6.92397 | 5.55754 |  |
|  | 4.41774 | 0.14217 | 1.50318 | 1.22386 | 0.42963 | 0 |
|  | 0.81319 | 0.36316 | 0 | 0 | 0 | 0 |
|  | 0.40667 | 0 | 0 | 0.09932 | 0 |  |
|  | 0 | 11.9556 | 27.2521 | 26.915 | 11.617 |  |
|  | 8.7491 | 3.3699 | 10.2461 | 9.66501 | 5.70182 |  |
|  | 2.42857 | 1.65195 | 1.89138 | 1.44948 | 0.9786 |  |
|  | 0.12415 | 0.9035 | 1.85437 | 2.84543 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 |  |


| 1 | 0 | 3.3789 | 24.1601 | 18.2609 | 15.6175 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 6.4364 | 2.5672 | 2.8808 | 1.45679 | 4.30936 |  |
|  | 5.37546 | 0.83462 | 1.0417 | 0.17687 | 0.93002 | 0 |
|  | 1.16367 | 1.35555 | 2.13469 | 0 | 0 | 0 |
|  | 0 | 0 | 0.10455 | 1.05699 | 1.05699 |  |
| 1 | 0 | 21.1326 | 26.0624 | 27.4407 | 19.3966 |  |
|  | 11.162 | 11.8984 | 2.0858 | 1.94805 | 2.06037 |  |
|  | 1.40477 | 1.26444 | 1.18208 | 1.11282 | 0.60453 | 0 |
|  | 0.56543 | 0 | 1.65612 | 0 | 1.11282 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 13.6486 | 24.6563 | 23.7656 | 19.809 |  |
|  | 9.1065 | 4.0748 | 7.1824 | 2.24506 | 0.30904 |  |
|  | 1.9532 | 1.05653 | 0.3958 | 0.48413 | 0.15127 |  |
|  | 0.09459 | 0.41663 | 0 | 0.15892 | 0 | 0 |
|  | 0.09459 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 2.3036 | 23.2228 | 26.7927 | 11.0111 |  |
|  | 9.7258 | 11.4579 | 5.9073 | 3.97145 | 0.13376 |  |
|  | 1.82684 | 2.35364 | 0.36751 | 2.05882 | 0.09932 |  |
|  | 0.67124 | 0 | 0.09932 | 1.10408 | 0 | 0 |
|  | 0 | 1.00247 | 0 | 0 | 0 | 0 |
| 1 | 0 | 3.7142 | 12.4853 | 23.6131 | 21.5683 |  |
|  | 14.7024 | 11.8911 | 8.5158 | 7.77601 | 6.54977 |  |
|  | 1.0211 | 6.06254 | 1.10408 | 4.6406 | 1.12071 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
|  | 2.42532 | 1.00247 | 0 | 0 | 0 | 0 |
| 1 | 1.2565 | 5.2342 | 25.2683 | 31.1232 | 13.363 |  |
|  | 19.2418 | 13.2925 | 24.9744 | 7.5189 | 2.67556 |  |
|  | 3.84886 | 1.54683 | 1.32124 | 1.05685 | 1.5451 |  |
|  | 1.02614 | 1.23376 | 0 | 0 | 1.02614 |  |
|  | 0.21026 | 0 | 0 | 0 | 0 | 0 |
|  | 0 |  |  |  |  |  |
| 1 | 0.56543 | 5.0564 | 10.4716 | 13.1777 | 16.4052 |  |
|  | 13.1156 | 12.5791 | 7.5394 | 7.55054 | 3.25374 |  |
|  | 3.63526 | 1.02121 | 2.84755 | 3.7998 | 0 |  |
|  | 3.83094 | 0.3368 | 1.03257 | 0.18197 | 0.2225 |  |
|  | 1.67078 | 0 | 0 | 0 | 0 | 0 |
|  | 0 |  |  |  |  |  |
| 1 | 0.20429 | 14.2613 | 29.7948 | 14.0505 | 14.3579 |  |
|  | 10.8978 | 9.6971 | 12.9744 | 2.26091 | 2.49797 |  |
|  | 4.98397 | 2.59738 | 1.21161 | 0.08277 | 0 |  |
|  | 0.37593 | 0 | 0 | 0 | 1.20179 | 0 |
|  | 0.17698 | 0 | 0 | 0 | 0 | 0 |
| 1 | 1.41071 | 5.7822 | 20.1244 | 19.6491 | 11.0867 |  |
|  | 8.3659 | 7.942 | 2.5658 | 2.48221 | 2.40893 |  |
|  | 1.51529 | 5.46344 | 1.19731 | 1.9858 | 0.15863 |  |
|  | 0.29223 | 0 | 0 | 1.98612 | 0.3958 |  |
|  | 0.19391 | 0 | 0 | 0 | 0 | 0 |
|  | 0 |  |  |  |  |  |
| 1 | 0 | 11.7026 | 23.6762 | 28.3671 | 21.9248 |  |
|  | 7.6216 | 11.5812 | 6.8262 | 6.37338 | 5.55139 |  |
|  | 1.41046 | 3.92975 | 4.62352 | 5.81395 | 0.83888 | 0 |
|  | 0 | 0 | 0 | 0.49328 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |

FSP-UK
20032018
110.70 .75

127

| 1 | 0.000374783 | 0.16425357 | 0.333157743 | 0.342104285 | 0.307789686 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.027687761 | 0.043434988 | 0.001216353 | 0.060658762 | 0.045237007 |  |
|  | 0.076077736 | 0.004223941 | 0.004487331 | 0.001602123 | 0.000293308 |  |
|  | 0.001509661 | 0 | 0 | 0.000146654 | 0.000146654 | 0 |
|  | 0.000293308 | 0.000146654 | 0 | 0.000146654 | 0 | 0 |
| 1 | 0.000167383 | 0.152985608 | 0.54713534 | 0.313153971 | 0.258367807 |  |
|  | 0.125306132 | 0.057695467 | 0.087452431 | 0.035445215 | 0.015927246 |  |
|  | 0.016567547 | 0.010042336 | 0.006018962 | 0.006226747 | 0.004546988 |  |
|  | 0.001033787 | 0.001162246 | 0.000662934 | 0.003393134 | 0.001437831 |  |
|  | 0.000199344 | 0.001241386 | 0.0022483 | 0.000732909 | 0.000824522 | 0 |
|  | 0 |  |  |  |  |  |
| 1 | 0 | 0.103329518 | 0.19641048 | 0.241991372 | 0.109126628 |  |
|  | 0.156802612 | 0.145326301 | 0.036140277 | 0.029396359 | 0.014350801 |  |
|  | 0.015371889 | 0.007192957 | 0.006752774 | 0.001868139 | 0.009940521 |  |
|  | 0.00740716 | 0.002378835 | 0.002716705 | 0.002140931 | 0.001742275 |  |
|  | 0.000590406 | 0.003395581 | 0.000675262 | 0 | 0 |  |
|  | 0.00023526 | 0 |  |  |  |  |
| 1 | 0.00361101 | 0.153691116 | 0.340745611 | 0.155260454 | 0.213275765 |  |
|  | 0.09839317 | 0.115716826 | 0.133528151 | 0.026403531 | 0.025886412 |  |
|  | 0.018344075 | 0.013299442 | 0.009312048 | 0.001825551 | 0.004269052 |  |
|  | 0.003885913 | 0.003547444 | 0.002248216 | 0.002021527 | 0.001306935 |  |
|  | 0.000573667 | 0.000776348 | 0.000582087 | 0.000884103 | $3.84393 \mathrm{e}-05$ |  |
|  | $3.84393 \mathrm{e}-05$ | 0 |  |  |  |  |
| 1 | 0.000949919 | 0.119241548 | 0.44701361 | 0.204189719 | 0.077363475 |  |
|  | 0.090584633 | 0.059564942 | 0.048392134 | 0.103423228 | 0.018747854 |  |
|  | 0.026135604 | 0.00518708 | 0.014899006 | 0.004306601 | 0.004122799 |  |
|  | 0.003789578 | 0 | 0.000313876 | 0.000511068 | 0.000980892 |  |
|  | 0.00080048 | 0.001114356 | 0.00167399 | 0.000972199 | 0 | 0 |
|  | 0 |  |  |  |  |  |
| 1 | $2.92679 \mathrm{e}-05$ | 0.21902938 | 0.304310597 | 0.264563006 | 0.247311278 |  |
|  | 0.043037336 | 0.037404414 | 0.014603872 | 0.056648435 | 0.032857499 |  |
|  | 0.002040635 | 0.010387516 | 0.005144875 | 0.000344659 | 0.001847508 |  |
|  | 0.001254609 | 0.000256856 | 0.002322059 | 0.001753791 | 0.000123486 |  |
|  | 0.001947411 | 0.00147103 | $4.7289 \mathrm{e}-05$ | 0.000111018 | 0.001606173 |  |
|  | $1.16069 \mathrm{e}-05$ | 0 |  |  |  |  |
| 1 | 0 | 0.087175684 | 0.299624141 | 0.311159869 | 0.161288882 |  |
|  | 0.060718142 | 0.039957338 | 0.028000462 | 0.015193089 | 0.017913114 |  |
|  | 0.047375509 | 0.007065787 | 0.002906977 | 0.002808564 | 0.003424814 | 0 |
|  | 0.002300992 | 0 | 0 | 0 | 0.001448773 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.119863413 | 0.196874246 | 0.245797705 | 0.181168944 |  |
|  | 0.127269974 | 0.035676999 | 0.020992322 | 0.027191027 | 0.017568869 |  |
|  | 0.023533383 | 0.011131766 | 0.004017553 | 0.002867057 | 0.009837834 |  |
|  | 0.006157131 | 0 | 0 | 0.001716561 | 0 |  |
|  | 0.00143812 | 0.001962611 | 0.00143812 | 0 | 0 | 0 |
|  | 0 |  |  |  |  |  |
| 1 | 0 | 0.08434561 | 0.454242063 | 0.099822858 | 0.198143553 |  |
|  | 0.092413349 | 0.051026632 | 0.004545029 | 0.013054823 | 0.007279282 |  |
|  | 0.010694232 | 0.012408527 | 0.013283726 | 0.001237655 | 0.003758948 |  |



| 3.33 | 569.33 | 159.31 | 112.20 | 42.39 | 44.18 | 21.30 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 30.70 | 7.94 | 5.60 | 5.48 | 5.88 | 5.21 |
| 4.11 | 276.52 | 436.07 | 135.24 | 82.61 | 58.75 | 29.82 |
|  | 23.11 | 22.81 | 11.35 | 3.31 | 8.58 | 5.80 |
| 6.81 | 347.00 | 282.99 | 271.57 | 54.29 | 49.16 | 24.17 |
|  | 27.27 | 20.69 | 23.17 | 11.03 | 8.54 | 4.49 |
| 6.93 | 139.39 | 287.26 | 193.06 | 187.53 | 57.49 | 45.54 |
|  | 26.86 | 14.72 | 8.08 | 17.93 | 7.45 | 5.17 |
| 7.59 | 146.04 | 118.70 | 100.89 | 81.14 | 87.63 | 23.24 |
|  | 21.23 | 16.83 | 12.69 | 13.77 | 12.60 | 5.11 |
| 7.37 | 300.18 | 244.82 | 114.67 | 60.06 | 66.02 | 58.33 |
|  | 14.54 | 6.74 | 13.71 | 5.51 | 6.41 | 4.75 |
| 7.30 | 188.05 | 166.31 | 103.86 | 61.72 | 44.52 | 23.65 |
|  | 35.65 | 9.80 | 9.76 | 8.10 | 8.57 | 3.78 |
| 7.03 | 264.75 | 137.13 | 101.88 | 64.10 | 27.00 | 25.49 |
|  | 13.29 | 26.52 | 5.87 | 9.91 | 2.81 | 2.98 |
| 8.15 | 194.23 | 235.47 | 112.00 | 69.45 | 33.41 | 16.90 |
|  | 19.70 | 14.88 | 26.19 | 2.84 | 4.35 | 1.86 |
| 9.62 | 400.24 | 142.06 | 135.26 | 69.22 | 46.01 | 25.81 |
|  | 13.47 | 11.17 | 10.68 | 12.43 | 4.64 | 3.50 |
| 9.44 | 280.20 | 169.83 | 62.21 | 62.54 | 27.88 | 19.67 |
|  | 8.64 | 3.97 | 4.69 | 2.63 | 4.92 | 2.28 |
| UK-WEC-BTS |  |  |  |  |  |  |
| 19882013 |  |  |  |  |  |  |
| 110.750 .8 |  |  |  |  |  |  |
| 19 |  |  |  |  |  |  |
| 128.20 | 2.00 | 39.00 | 129.00 | 52.00 | 75.00 | 22.00 |
|  | 0.00 | 12.00 | 3.00 |  |  |  |
| 165.70 | 5.00 | 56.00 | 120.00 | 107.00 | 34.00 | 40.00 |
|  | 17.00 | 5.00 | 7.00 |  |  |  |
| 175.70 | 23.00 | 52.00 | 76.00 | 31.00 | 24.00 | 7.00 |
|  | 15.00 | 3.00 | 6.00 |  |  |  |
| 171.70 | 11.00 | 231.00 | 79.00 | 51.00 | 23.00 | 21.00 |
|  | 5.00 | 17.00 | 4.00 |  |  |  |
| 196.60 | 5.00 | 140.00 | 316.00 | 44.00 | 36.00 | 12.00 |
|  | 7.00 | 5.00 | 11.00 |  |  |  |
| 189.20 | 5.00 | 54.00 | 115.00 | 105.00 | 14.00 | 10.00 |
|  | 9.00 | 3.00 | 3.00 |  |  |  |
| 205.90 | 6.00 | 47.00 | 106.00 | 62.00 | 44.00 | 5.00 |
|  | 5.00 | 2.00 | 3.00 |  |  |  |
| 187.20 | 14.00 | 37.00 | 44.00 | 42.00 | 26.00 | 31.00 |
|  | 4.00 | 5.00 | 5.00 |  |  |  |
| 184.40 | 28.00 | 112.00 | 67.00 | 25.00 | 32.00 | 20.00 |
|  | 17.00 | 3.00 | 2.00 |  |  |  |
| 184.70 | 11.00 | 130.00 | 126.00 | 43.00 | 14.00 | 16.00 |
|  | 13.00 | 14.00 | 5.00 |  |  |  |
| 185.50 | 11.00 | 141.00 | 114.00 | 76.00 | 22.00 | 10.00 |
|  | 14.00 | 6.00 | 8.00 |  |  |  |
| 187.90 | 11.00 | 97.00 | 128.00 | 47.00 | 23.00 | 8.00 |
|  | 4.00 | 4.00 | 4.00 |  |  |  |
| 180.40 | 12.00 | 136.00 | 70.00 | 52.00 | 23.00 | 16.00 |
|  | 5.00 | 3.00 | 5.00 |  |  |  |


| 178.00 | 9.00 | 197.00 | 162.00 | 52.00 | 31.00 | 12.00 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 12.00 | 4.00 | 1.00 |  |  |  |
| 180.00 | 6.00 | 37.00 | 113.00 | 48.00 | 27.00 | 6.00 |
|  | 3.00 | 2.00 | 0.00 |  |  |  |
| 170.70 | 23.00 | 124.00 | 78.00 | 56.00 | 28.00 | 6.00 |
|  | 1.00 | 1.00 | 2.00 |  |  |  |
| 164.90 | 16.00 | 110.00 | 120.00 | 24.00 | 15.00 | 10.00 |
|  | 16.00 | 9.00 | 4.00 |  |  |  |
| 186.60 | 8.00 | 110.00 | 39.00 | 53.00 | 12.00 | 12.00 |
|  | 6.00 | 2.00 | 4.00 |  |  |  |
| 184.70 | 5.00 | 120.00 | 95.00 | 26.00 | 37.00 | 10.00 |
|  | 7.00 | 9.00 | 0.00 |  |  |  |
| 181.00 | 7.00 | 188.00 | 135.00 | 50.00 | 11.00 | 23.00 |
|  | 3.00 | 3.00 | 1.00 |  |  |  |
| 174.70 | 10.00 | 85.00 | 158.00 | 77.00 | 40.00 | 2.00 |
|  | 14.00 | 3.00 | 6.00 |  |  |  |
| 172.00 | 11.00 | 104.00 | 126.00 | 96.00 | 49.00 | 13.00 |
|  | 13.00 | 12.00 | 1.00 |  |  |  |
| 179.90 | 20.00 | 175.00 | 154.00 | 84.00 | 59.00 | 31.00 |
|  | 20.00 | 7.00 | 12.00 |  |  |  |
| $176.20$ | 9.00 | 156.00 | 231.00 | 62.00 | 39.00 | 25.00 |
|  | 24.00 | 8.00 | 2.00 |  |  |  |
| 179.70 | 3.00 | 47.00 | 162.00 | 125.00 | 40.00 | 27.00 |
|  | 13.00 | 3.00 | 6.00 |  |  |  |
| 181.60 | 4.00 | 36.00 | 100.00 | 106.00 | 80.00 | 21.00 |
|  | 9.00 | 6.00 | 3.00 |  |  |  |

Table 31.7. Sole in Division 7.e. Detailed XSA survivor diagnostics.

FLR XSA Diagnostics 2019-05-10 11:00:52

CPUE data from indices

Catch data for 50 years 1969 to 2018. Ages 2 to 12.
fleet first age last age first year last year alpha beta
1 UK-CBT-late 31120032018 <NA> <NA $>$
2 UK-COT $3 \quad 11 \quad 1988 \quad 2015<N A><N A>$
3 Q1SWBeam-nonoffset 21120062018 <NA> <NA>
4 FSP-UK 21120042018 <NA> <NA>

Time series weights :

Tapered time weighting applied
Power $=3$ over 15 years
Catchability analysis :
Catchability independent of size for all ages
Catchability independent of age for ages > 7
Terminal population estimation :
Survivor estimates shrunk towards the mean F of the final 3 years or the 5 oldest ages.
S.E. of the mean to which the estimates are shrunk $=0.5$

Minimum standard error for population
estimates derived from each fleet $=0.4$
prior weighting not applied
Regression weights
year
age 2009201020112012201320142015201620172018 all 0.4820 .610 .7250 .820 .8930 .9440 .9760 .9930 .9991


XSA population number (Thousand)
age
$\begin{array}{llllllllllll}\text { year } & 2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10 & 11 & 12\end{array}$
20093943373824181713926398414213294253467 201050383410286818221233645298304157218507 201137254493275422201336922463223208122404 201235973283359120391624987685343164155500 2013326032252755274914641132673490246107518

2014321528362610203319211058821469350177633 2015435427212262191914351349797636350259705 20165735377121491671141610501024612501274889 2017406151023013149912491044720747491406243 201850183557348724091064867821562637410284

Estimated population abundance at 1st Jan 2019 age
year $2 \begin{array}{lllllllllll}3 & 4 & 5 & 6 & 7 & 8 & 9 & 10 & 11 & 12\end{array}$
201904464279427651920640636513415498161

Fleet: UK-CBT-late

Log catchability residuals.
year
age 20032004200520062007200820092010201120122013 3-0.264-0.472 0.5920 .7710 .8930 .542 0.218-0.608 $0.123-0.335-0.648$ $4-0.554-1.197 \quad 0.239 \quad 0.3130 .2950 .2510 .046-0.182-0.422-0.0920 .063$ $5-0.766-1.347-0.025 \quad 0.365 \quad 0.172-0.0350 .156 \quad 0.088-0.051-0.3690 .188$ $6-1.646-0.949 \quad 0.114-0.050-0.006 \quad 0.1220 .186-0.040-0.044-0.162-0.030$ $7-1.262-1.9190 .1610 .2340 .040 \quad 0.078-0.1950 .005-0.028-0.060-0.095$ 8 -1.351-1.129-0.604 0.368 0.098 $0.292-0.156 \quad 0.036-0.169-0.040 \quad 0.189$ $9-0.667-1.3500 .143$ 0.100 $0.047-0.006-0.237-0.053-0.247-0.1250 .002$ $10-1.174-1.0290 .192 \quad 0.156-0.359 \quad 0.048-0.369-0.4640 .0650 .1180 .099$ 11-1.280-1.157 0.360 0.100-0.003-0.002-0.458-0.071 0.008-0.021 0.116 year
age 20142015201620172018
$3-0.200-0.0090 .2190 .4460 .065$
4 -0.067-0.106 0.0780 .0710 .242
$5-0.046 \quad 0.090-0.095 \quad 0.059-0.025$
6 0.117-0.064 0.029-0.037 0.052
$\begin{array}{lllllllllll}7 & 0.025 & -0.121 & 0.098 & -0.039 & 0.225\end{array}$
8 -0.127-0.179-0.057-0.050 -0.079
9 0.019-0.323-0.321-0.243-0.176
$10-0.043-0.092-0.207-0.360-0.372$
$11-0.025-0.231-0.345-0.4240 .005$

Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

$$
\begin{array}{lllllllll}
3 & 4 & 5 & 6 & 7 & 8 & 9 & 10 & 11
\end{array}
$$

Mean_Logq -5.0907-4.5145-4.4365-4.3698-4.4145-4.4145-4.4145-4.4145-4.4145
S.E_Logq $0.44820 .44820 .44820 .44820 .44820 .4482 \quad 0.44820 .44820 .4482$

Fleet: UK-COT
Log catchability residuals.
year
age $\begin{array}{lllllllllll}1988 & 1989 & 1990 & 1991 & 1992 & 1993 & 1994 & 1995 & 1996 & 1997 & 1998\end{array}$
31.0080 .6470 .9190 .6450 .6100 .5680 .8010 .2440 .6841 .0120 .672 $\begin{array}{lllllllllllllllllll}4 & 0.651 & 0.467 & 0.464 & 0.010 & -0.033 & 0.203 & 0.457 & 0.539 & 0.525 & 0.333 & 0.337\end{array}$
$5 \quad 0.5840 .464 \quad 0.349-0.213-0.3720 .3170 .2440 .2340 .3530 .3930 .064$ $6 \quad 0.4550 .4120 .572-0.459-1.241-0.005-0.018-0.1430 .1170 .0840 .256$ 7 -0.144 $0.3740 .284-0.624-0.778 \quad 0.0060 .2940 .218-0.008 \quad 0.4220 .104$ $8 \quad 0.0470 .105 \quad 0.096-0.580-0.996-0.265-0.298 \quad 0.146$ 0.183-0.174-0.195 $9-0.3350 .348$ 0.312-0.941-1.099 0.211 0.219-0.299 0.115-0.015-0.295 10-1.061 0.059 0.129-0.655-1.528 -0.372 $0.804-0.399-0.040-0.721-0.244$ $11-0.1700 .623-0.236-0.911-1.8830 .069 \quad 0.285-0.1370 .010-0.3140 .151$ year
age 19992000200120022003200420052006200720082009 30.7710 .3460 .7060 .3970 .1260 .0441 .0290 .7250 .8040 .4680 .645 $4 \quad 0.2270 .306-0.044-0.385-0.103-0.5900 .7080 .2940 .2990 .1340 .211$ $50.3160 .1090 .150-0.404-0.271-0.7610 .3300 .298 \quad 0.204-0.1940 .151$

# 6 0.080-0.036 0.069 0.070 -0.895-0.345 $0.454-0.199 \quad 0.040-0.028 \quad 0.107$ <br> $7 \quad 0.379-0.3730 .212-0.015-0.496-1.3190 .5390 .090 \quad 0.1250 .051-0.199$ 8 0.618-0.195-0.170-0.120 -0.558-0.591-0.189 $0.308 \quad 0.2170 .2470 .030$ $9-0.0150 .472$ 0.364-0.883-0.048 -0.762 $0.5790 .0250 .1320 .017-0.060$ $100.1320 .088 \quad 0.554-0.309-0.544-0.439 \quad 0.559 \quad 0.009-0.207-0.088-0.100$ $110.4680 .020 \quad 0.288 \quad 0.482-0.267-0.517 \quad 0.932-0.074-0.068 \quad 0.169-0.114$ year <br> age 201020112012201320142015 <br> $3-0.547$ 0.195-0.013-0.459-0.168-0.009 <br> $4-0.086-0.4540 .2110 .059-0.031-0.092$ <br> 5 $0.090-0.168-0.122 \quad 0.208-0.033-0.083$ <br> $6-0.021-0.206 \quad 0.058-0.013 \quad 0.117-0.029$ <br> $7 \quad 0.092-0.139 \quad 0.103-0.077 \quad 0.103-0.044$ <br> $8 \quad 0.146-0.4870 .1780 .3060 .010-0.180$ <br> $9 \quad 0.078-0.3810 .1050 .097 \quad 0.228-0.303$ <br> $10-0.226-0.1410 .3710 .3200 .2450 .123$ <br> $110.149-0.238 \quad 0.351 \quad 0.395 \quad 0.222-0.100$ 

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

$$
\begin{array}{llllllll}
3 & 4 & 5 & 6 & 7 & 8 & 9 & 10
\end{array}
$$

Mean_Logq -14.1872 -13.6861-13.5925-13.5089-13.6137-13.6137-13.6137-13.6137 S.E_Logq $\quad 0.4363 \quad 0.4363 \quad 0.4363 \quad 0.4363 \quad 0.4363 \quad 0.4363 \quad 0.4363 \quad 0.4363$

## 11

Mean_Logq -13.6137
S.E_Logq 0.4363

Fleet: Q1SWBeam-nonoffset
Log catchability residuals.
year
age 20062007200820092010201120122013201420152016
$20.5660 .5800 .460-0.6920 .8910 .758-0.989-0.408-0.046-0.388$
$3-0.1220 .5110 .2960 .0830 .242-0.088 \quad 0.157-0.439 \quad 0.397-0.4410 .277$
$4-0.199-0.100 \quad 0.294-0.0430 .189 \quad 0.093-0.058$ 0.087 0.418 -0.299-0.174
$5 \quad 0.337-1.308 \quad 0.0730 .1630 .3150 .138-0.3610 .018-0.1610 .0940 .097$
$6 \quad 0.520-0.1620 .583-0.151 \quad 0.101-0.180-0.300 \quad 0.211 \quad 0.2130 .115-0.059$
7 0.193-0.704-0.573-0.382 $0.677-0.758$ 0.223 $0.1120 .284-0.015-0.007$
8 0.408-0.683 1.180-0.302-0.299 $0.499-0.0830 .3051 .164-0.0030 .298$
$9 \quad 0.4050 .389 \quad 0.943-0.318-0.3720 .068 \quad 0.2110 .5270 .5290 .220-0.951$
$100.6450 .9640 .5500 .4410 .323-1.848-2.4271 .043-0.210-0.023-0.652$
$\begin{array}{llllllllllllllllllll}11 & 1.794 & 1.417 & 0.534 & 0.805 & -0.387 & 0.539 & 0.239 & 0.021 & 0.833 & 0.387 & 0.646\end{array}$
year
age 20172018
2 -0.186 0.304
3-0.391 0.108
$4-0.2000 .022$
$5-0.0440 .143$
$6-0.187-0.095$
$7-0.2250 .350$
8-0.982-0.096
$9-1.0670 .185$
10-0.674-0.088
11-0.889-0.897

Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

$$
\begin{array}{lllllllll}
2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10
\end{array}
$$

Mean_Logq -6.3448-5.0782-4.7932-4.8029-4.7550 -4.6117-4.6117-4.6117-4.6117 S.E_Logq 0.57830 .57830 .57830 .57830 .57830 .57830 .57830 .57830 .5783 11
Mean_Logq-4.6117
S.E_Logq 0.5783

## Fleet: FSP-UK

Log catchability residuals.
year
age 20042005200620072008200920102011201220132014
$20.8850 .0890 .3780 .229 \quad 0.756-0.067-0.016-0.057-0.635-0.4410 .283$
3 $0.368-0.0280 .1150 .290-0.017-0.177-0.5430 .0250 .088 \quad 0.1110 .115$
$4 \quad 0.401-0.1240 .061-0.041 \quad 0.054 \quad 0.192-0.234-1.062-0.028 \quad 0.4010 .397$
5 $0.315-0.1660 .287-0.127 \quad 0.634-0.054-0.012-0.119-0.1660 .3010 .471$
$6-0.138 \quad 0.3570 .316 \quad 0.079-0.148-0.267 \quad 0.135-0.256-0.170 \quad 0.2770 .259$
$70.1370 .6370 .6700 .498-0.136 \quad 0.304-0.261-0.286-0.7120 .1110 .423$
$81.0800 .0430 .9560 .179-0.418-0.077-0.049-2.013-0.1000 .0990 .127$

100.4180 .1150 .7680 .178 0.261-0.190 $0.383-0.748$-1.268 -1.023 0.509
$110.6450 .8120 .7591 .184-1.6310 .9040 .365 \quad 0.227-0.5140 .002-1.884$
year
age 2015201620172018
2 0.065-0.260 0.2850 .271
$3-0.0930 .2190 .085-0.166$
4 0.377-0.342-0.010 0.078
$5-0.047-0.168-0.428-0.039$
6 0.005-0.027-0.218 0.102
$\begin{array}{lllll}7 & 0.077 & 0.067-0.098 & 0.093\end{array}$
80.247 0.318-0.341 0.547

9 0.176-0.459-0.031-0.419
$10-0.1150 .196-0.6740 .226$
$110.234-0.363-0.7350 .665$

Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

$$
\begin{array}{lllllllll}
2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10
\end{array}
$$

Mean_Logq -10.5476 -9.0621-8.9453 -8.9788 -9.1030 -9.3004-9.3004-9.3004-9.3004 S.E_Logq 0.48630 .48630 .48630 .48630 .48630 .48630 .48630 .48630 .4863 11
Mean_Logq -9.3004
S.E_Logq 0.4863

Terminal year survivor and F summaries:
,Age 2 Year class $=2016$
source
scaledWts survivors yrcls
Q1SWBeam-nonoffset 0.20960532016 FSP-UK $\quad 0.479 \quad 58522016$
fshk $\quad 0.312 \quad 24012016$
,Age 3 Year class =2015
source
scaledWts survivors yrcls
UK-CBT-late $0.238 \quad 29832015$
Q1SWBeam-nonoffset $0.278 \quad 31132015$
FSP-UK $\quad 0.278 \quad 23672015$
fshk $\quad 0.205 \quad 21452015$
,Age 4 Year class =2014
source
scaledWts survivors yrcls
UK-CBT-late $\quad 0.281 \quad 35212014$
Q1SWBeam-nonoffset $0.281 \quad 28272014$
FSP-UK $\quad 0.233 \quad 29892014$

```
fshk 0.205 1801 2014
,Age 5 Year class=2013
ource
    scaledWts survivors yrcls
UK-CBT-late 0.268 1872 2013
Q1SWBeam-nonoffset 0.268 2216 2013
FSP-UK 0.268 1846 2013
fshk 0.195 1092 2013
,Age 6 Year class =2012
source
    scaledWts survivors yrcls
UK-CBT-late 0.252 675 2012
Q1SWBeam-nonoffset 0.252 582 2012
FSP-UK 0.252 709 2012
fshk 0.243 1261 2012
,Age 7 Year class =2011
source
    scaledWts survivors yrcls
UK-CBT-late 0.271 796 2011
Q1SWBeam-nonoffset 0.245 902 2011
FSP-UK 0.271 698 2011
fshk 0.214 681 2011
,Age 8 Year class =2010
source
scaledWts survivors yrcls
UK-CBT-late \(0.385 \quad 4742010\)
Q1SWBeam-nonoffset \(0.136 \quad 4662010\)
FSP-UK 0.1228872010
fshk \(\quad 0.357 \quad 11942010\)
,Age 9 Year class =2009
source
scaledWts survivors yrcls
UK-CBT-late \(0.317 \quad 3482009\)
Q1SWBeam-nonoffset \(0.117 \quad 4992009\)
FSP-UK \(\quad 0.317 \quad 2732009\)
fshk \(\quad 0.249 \quad 8322009\)
```

,Age 10 Year class $=2008$
source

> scaledWts survivors yrcls

UK-CBT-late $\quad 0.458 \quad 3432008$
Q1SWBeam-nonoffset $0.055 \quad 4562008$
FSP-UK $\quad 0.147 \quad 6252008$
fshk $0.339 \quad 6582008$
,Age 11 Year class =2007
source
scaledWts survivors yrcls
UK-CBT-late $\quad 0.345 \quad 1612007$
Q1SWBeam-nonoffset $0.092 \quad 662007$
FSP-UK $\quad 0.054 \quad 3132007$
$\begin{array}{lll}\text { fshk } & 0.509 \quad 6832007\end{array}$

Table 31.8. Sole in Division 7.e. Estimated stock numbers-at-age (thousands).

| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 1874 | 2380 | 625 | 966 | 1513 | 159 | 507 | 572 | 262 | 90 | 636 | 9585 |
| 1970 | 1343 | 1611 | 1848 | 490 | 732 | 1170 | 124 | 412 | 494 | 218 | 1123 | 9564 |
| 1971 | 3826 | 1164 | 1237 | 1365 | 358 | 584 | 952 | 100 | 340 | 397 | 821 | 11144 |
| 1972 | 2568 | 3414 | 863 | 885 | 1047 | 262 | 452 | 713 | 81 | 274 | 542 | 11102 |
| 1973 | 2264 | 2185 | 2698 | 621 | 691 | 840 | 224 | 386 | 518 | 37 | 1222 | 11687 |
| 1974 | 3107 | 1981 | 1600 | 2029 | 478 | 532 | 646 | 187 | 300 | 440 | 850 | 12150 |
| 1975 | 2967 | 2769 | 1461 | 1238 | 1667 | 365 | 406 | 544 | 138 | 248 | 1756 | 13559 |
| 1976 | 2792 | 2607 | 1966 | 1160 | 931 | 1399 | 304 | 317 | 468 | 105 | 1598 | 13645 |
| 1977 | 6557 | 2367 | 1960 | 1330 | 897 | 714 | 1178 | 230 | 231 | 375 | 1866 | 17703 |
| 1978 | 4658 | 5527 | 1839 | 1408 | 1007 | 714 | 580 | 995 | 199 | 186 | 1385 | 18498 |
| 1979 | 4389 | 3976 | 3933 | 1334 | 1070 | 732 | 548 | 456 | 827 | 144 | 1493 | 18902 |
| 1980 | 4703 | 3755 | 2834 | 2787 | 970 | 751 | 497 | 397 | 327 | 650 | 1702 | 19374 |
| 1981 | 8131 | 4089 | 2866 | 2092 | 1923 | 758 | 506 | 316 | 298 | 243 | 934 | 22156 |
| 1982 | 4680 | 7124 | 2933 | 1974 | 1448 | 1370 | 516 | 337 | 214 | 214 | 1035 | 21846 |
| 1983 | 3867 | 4113 | 5066 | 1782 | 1260 | 976 | 1011 | 337 | 198 | 117 | 828 | 19557 |
| 1984 | 5969 | 3413 | 3006 | 3088 | 1058 | 806 | 629 | 635 | 192 | 110 | 982 | 19888 |
| 1985 | 6984 | 5084 | 2457 | 1934 | 2073 | 649 | 535 | 446 | 430 | 123 | 532 | 21247 |
| 1986 | 3766 | 6046 | 2982 | 1504 | 1304 | 1320 | 417 | 392 | 306 | 309 | 529 | 18876 |
| 1987 | 5850 | 3174 | 3932 | 1775 | 961 | 874 | 874 | 298 | 283 | 191 | 755 | 18967 |
| 1988 | 3881 | 4830 | 2103 | 2520 | 1200 | 676 | 578 | 573 | 224 | 208 | 713 | 17507 |
| 1989 | 3738 | 3091 | 3002 | 1336 | 1588 | 730 | 466 | 369 | 415 | 166 | 744 | 15645 |
| 1990 | 2820 | 3012 | 1968 | 1544 | 736 | 953 | 446 | 317 | 233 | 268 | 740 | 13037 |
| 1991 | 7176 | 2227 | 1867 | 1227 | 871 | 434 | 611 | 268 | 189 | 139 | 657 | 15667 |
| 1992 | 3910 | 6065 | 1621 | 1230 | 836 | 579 | 305 | 447 | 178 | 136 | 529 | 15836 |
| 1993 | 3356 | 3238 | 4124 | 1070 | 831 | 647 | 417 | 218 | 335 | 136 | 345 | 14716 |
| 1994 | 2386 | 2837 | 2261 | 2678 | 635 | 543 | 442 | 303 | 140 | 250 | 489 | 12963 |
| 1995 | 3469 | 2066 | 1943 | 1515 | 1893 | 468 | 391 | 353 | 219 | 85 | 649 | 13049 |
| 1996 | 3966 | 3048 | 1576 | 1159 | 964 | 1321 | 299 | 258 | 261 | 167 | 653 | 13674 |


| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 3374 | 3242 | 2335 | 1080 | 766 | 649 | 951 | 206 | 175 | 190 | 415 | 13382 |
| 1998 | 4450 | 2847 | 2142 | 1424 | 668 | 522 | 403 | 696 | 145 | 139 | 654 | 14091 |
| 1999 | 3616 | 3774 | 2000 | 1428 | 969 | 405 | 328 | 288 | 509 | 98 | 484 | 13900 |
| 2000 | 6647 | 3005 | 2545 | 1334 | 914 | 634 | 258 | 200 | 209 | 358 | 372 | 16477 |
| 2001 | 5476 | 5722 | 2150 | 1589 | 858 | 609 | 472 | 183 | 116 | 141 | 529 | 17845 |
| 2002 | 3864 | 4817 | 3845 | 1440 | 965 | 522 | 382 | 331 | 113 | 64 | 297 | 16641 |
| 2003 | 5456 | 3180 | 3169 | 2678 | 935 | 567 | 350 | 245 | 210 | 72 | 317 | 17179 |
| 2004 | 2896 | 4368 | 2083 | 1960 | 1809 | 722 | 442 | 269 | 167 | 130 | 314 | 15161 |
| 2005 | 4068 | 2242 | 2925 | 1459 | 1351 | 1137 | 497 | 289 | 186 | 100 | 394 | 14647 |
| 2006 | 4698 | 3436 | 1583 | 1853 | 893 | 874 | 750 | 342 | 186 | 126 | 351 | 15091 |
| 2007 | 4039 | 3775 | 2361 | 983 | 1100 | 569 | 578 | 503 | 228 | 115 | 426 | 14680 |
| 2008 | 4426 | 3464 | 2606 | 1418 | 611 | 651 | 345 | 400 | 351 | 154 | 348 | 14773 |
| 2009 | 3943 | 3738 | 2418 | 1713 | 926 | 398 | 414 | 213 | 294 | 253 | 467 | 14778 |
| 2010 | 5038 | 3410 | 2868 | 1822 | 1233 | 645 | 298 | 304 | 157 | 218 | 507 | 16500 |
| 2011 | 3725 | 4493 | 2754 | 2220 | 1336 | 922 | 463 | 223 | 208 | 122 | 404 | 16870 |
| 2012 | 3597 | 3283 | 3591 | 2039 | 1624 | 987 | 685 | 343 | 164 | 155 | 500 | 16968 |
| 2013 | 3260 | 3225 | 2755 | 2749 | 1464 | 1132 | 673 | 490 | 246 | 107 | 518 | 16620 |
| 2014 | 3215 | 2836 | 2610 | 2033 | 1921 | 1058 | 821 | 469 | 350 | 177 | 633 | 16124 |
| 2015 | 4354 | 2721 | 2262 | 1919 | 1435 | 1349 | 797 | 636 | 350 | 259 | 705 | 16786 |
| 2016 | 5735 | 3771 | 2149 | 1671 | 1416 | 1050 | 1024 | 612 | 501 | 274 | 889 | 19093 |
| 2017 | 4061 | 5102 | 3013 | 1499 | 1249 | 1044 | 720 | 747 | 491 | 406 | 243 | 18575 |
| 2018 | 5018 | 3557 | 3487 | 2409 | 1064 | 867 | 821 | 562 | 637 | 410 | 284 | 19115 |

Table 31.9. Sole in Division 7.e. Estimated fishing mortality-at-age.

| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | Fbar(3-9) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 0.051 | 0.153 | 0.144 | 0.176 | 0.157 | 0.151 | 0.108 | 0.048 | 0.084 | 0.11 | 0.11 | 0.134 |
| 1970 | 0.043 | 0.164 | 0.202 | 0.213 | 0.126 | 0.106 | 0.115 | 0.093 | 0.118 | 0.112 | 0.112 | 0.146 |
| 1971 | 0.014 | 0.2 | 0.234 | 0.165 | 0.212 | 0.155 | 0.188 | 0.109 | 0.113 | 0.156 | 0.156 | 0.181 |
| 1972 | 0.062 | 0.136 | 0.228 | 0.147 | 0.12 | 0.059 | 0.059 | 0.219 | 0.69 | 0.23 | 0.23 | 0.138 |
| 1973 | 0.034 | 0.212 | 0.185 | 0.163 | 0.163 | 0.162 | 0.081 | 0.151 | 0.063 | 0.124 | 0.124 | 0.16 |
| 1974 | 0.015 | 0.205 | 0.156 | 0.097 | 0.17 | 0.171 | 0.072 | 0.199 | 0.089 | 0.14 | 0.14 | 0.153 |
| 1975 | 0.029 | 0.243 | 0.13 | 0.185 | 0.075 | 0.083 | 0.147 | 0.051 | 0.181 | 0.108 | 0.108 | 0.131 |
| 1976 | 0.065 | 0.185 | 0.291 | 0.158 | 0.166 | 0.072 | 0.176 | 0.216 | 0.122 | 0.151 | 0.151 | 0.18 |
| 1977 | 0.071 | 0.152 | 0.23 | 0.178 | 0.128 | 0.108 | 0.069 | 0.048 | 0.114 | 0.093 | 0.093 | 0.13 |
| 1978 | 0.058 | 0.24 | 0.221 | 0.174 | 0.22 | 0.165 | 0.14 | 0.085 | 0.226 | 0.167 | 0.167 | 0.178 |
| 1979 | 0.056 | 0.239 | 0.244 | 0.219 | 0.254 | 0.287 | 0.221 | 0.232 | 0.142 | 0.228 | 0.228 | 0.242 |
| 1980 | 0.04 | 0.17 | 0.204 | 0.271 | 0.147 | 0.295 | 0.352 | 0.188 | 0.198 | 0.236 | 0.236 | 0.232 |
| 1981 | 0.032 | 0.232 | 0.273 | 0.268 | 0.239 | 0.285 | 0.305 | 0.289 | 0.229 | 0.27 | 0.27 | 0.27 |
| 1982 | 0.029 | 0.241 | 0.398 | 0.349 | 0.295 | 0.203 | 0.325 | 0.433 | 0.503 | 0.353 | 0.353 | 0.321 |
| 1983 | 0.025 | 0.214 | 0.395 | 0.421 | 0.347 | 0.34 | 0.365 | 0.461 | 0.489 | 0.402 | 0.402 | 0.363 |
| 1984 | 0.06 | 0.229 | 0.341 | 0.298 | 0.389 | 0.309 | 0.244 | 0.29 | 0.35 | 0.317 | 0.317 | 0.3 |
| 1985 | 0.044 | 0.433 | 0.391 | 0.294 | 0.352 | 0.342 | 0.212 | 0.277 | 0.232 | 0.284 | 0.284 | 0.329 |
| 1986 | 0.071 | 0.33 | 0.419 | 0.348 | 0.3 | 0.312 | 0.238 | 0.226 | 0.37 | 0.29 | 0.29 | 0.31 |
| 1987 | 0.092 | 0.312 | 0.345 | 0.292 | 0.252 | 0.313 | 0.322 | 0.182 | 0.205 | 0.256 | 0.256 | 0.288 |
| 1988 | 0.128 | 0.375 | 0.354 | 0.362 | 0.397 | 0.272 | 0.348 | 0.223 | 0.201 | 0.289 | 0.289 | 0.333 |
| 1989 | 0.116 | 0.352 | 0.565 | 0.496 | 0.411 | 0.392 | 0.285 | 0.363 | 0.337 | 0.359 | 0.359 | 0.409 |
| 1990 | 0.136 | 0.378 | 0.372 | 0.472 | 0.428 | 0.345 | 0.41 | 0.416 | 0.413 | 0.404 | 0.404 | 0.403 |
| 1991 | 0.068 | 0.218 | 0.317 | 0.284 | 0.308 | 0.255 | 0.211 | 0.311 | 0.23 | 0.264 | 0.264 | 0.272 |
| 1992 | 0.089 | 0.286 | 0.315 | 0.293 | 0.156 | 0.228 | 0.236 | 0.19 | 0.166 | 0.196 | 0.196 | 0.244 |
| 1993 | 0.068 | 0.259 | 0.332 | 0.422 | 0.325 | 0.282 | 0.22 | 0.34 | 0.193 | 0.273 | 0.273 | 0.311 |
| 1994 | 0.044 | 0.279 | 0.301 | 0.247 | 0.205 | 0.229 | 0.125 | 0.222 | 0.403 | 0.238 | 0.238 | 0.23 |
| 1995 | 0.029 | 0.171 | 0.416 | 0.352 | 0.259 | 0.348 | 0.316 | 0.201 | 0.17 | 0.259 | 0.259 | 0.295 |
| 1996 | 0.102 | 0.167 | 0.278 | 0.315 | 0.296 | 0.229 | 0.273 | 0.286 | 0.22 | 0.261 | 0.261 | 0.263 |


| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | Fbar(3-9) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 0.07 | 0.314 | 0.394 | 0.381 | 0.284 | 0.378 | 0.212 | 0.252 | 0.131 | 0.252 | 0.252 | 0.316 |
| 1998 | 0.065 | 0.253 | 0.305 | 0.285 | 0.399 | 0.363 | 0.234 | 0.213 | 0.292 | 0.301 | 0.301 | 0.293 |
| 1999 | 0.085 | 0.294 | 0.305 | 0.347 | 0.324 | 0.352 | 0.398 | 0.22 | 0.251 | 0.31 | 0.31 | 0.32 |
| 2000 | 0.05 | 0.235 | 0.371 | 0.341 | 0.306 | 0.196 | 0.242 | 0.441 | 0.295 | 0.297 | 0.297 | 0.305 |
| 2001 | 0.028 | 0.298 | 0.301 | 0.399 | 0.397 | 0.366 | 0.254 | 0.38 | 0.492 | 0.379 | 0.379 | 0.342 |
| 2002 | 0.095 | 0.319 | 0.262 | 0.332 | 0.432 | 0.3 | 0.342 | 0.353 | 0.36 | 0.359 | 0.359 | 0.334 |
| 2003 | 0.122 | 0.323 | 0.38 | 0.292 | 0.159 | 0.149 | 0.162 | 0.285 | 0.379 | 0.227 | 0.227 | 0.25 |
| 2004 | 0.156 | 0.301 | 0.256 | 0.272 | 0.365 | 0.273 | 0.323 | 0.27 | 0.418 | 0.334 | 0.334 | 0.295 |
| 2005 | 0.069 | 0.248 | 0.357 | 0.391 | 0.335 | 0.316 | 0.274 | 0.343 | 0.292 | 0.296 | 0.296 | 0.323 |
| 2006 | 0.119 | 0.275 | 0.376 | 0.421 | 0.35 | 0.313 | 0.3 | 0.303 | 0.378 | 0.301 | 0.301 | 0.334 |
| 2007 | 0.054 | 0.271 | 0.41 | 0.375 | 0.425 | 0.401 | 0.269 | 0.261 | 0.295 | 0.278 | 0.278 | 0.344 |
| 2008 | 0.069 | 0.259 | 0.319 | 0.326 | 0.328 | 0.352 | 0.385 | 0.208 | 0.226 | 0.313 | 0.313 | 0.311 |
| 2009 | 0.045 | 0.165 | 0.183 | 0.229 | 0.261 | 0.191 | 0.209 | 0.204 | 0.197 | 0.158 | 0.158 | 0.206 |
| 2010 | 0.014 | 0.113 | 0.156 | 0.211 | 0.191 | 0.232 | 0.19 | 0.281 | 0.147 | 0.176 | 0.176 | 0.196 |
| 2011 | 0.026 | 0.124 | 0.201 | 0.213 | 0.203 | 0.196 | 0.2 | 0.204 | 0.191 | 0.276 | 0.276 | 0.192 |
| 2012 | 0.009 | 0.076 | 0.167 | 0.231 | 0.261 | 0.283 | 0.236 | 0.231 | 0.329 | 0.302 | 0.302 | 0.212 |
| 2013 | 0.039 | 0.112 | 0.204 | 0.258 | 0.225 | 0.221 | 0.26 | 0.235 | 0.231 | 0.252 | 0.252 | 0.216 |
| 2014 | 0.067 | 0.126 | 0.208 | 0.249 | 0.254 | 0.183 | 0.155 | 0.194 | 0.203 | 0.176 | 0.176 | 0.196 |
| 2015 | 0.044 | 0.136 | 0.203 | 0.204 | 0.212 | 0.175 | 0.164 | 0.139 | 0.144 | 0.131 | 0.131 | 0.176 |
| 2016 | 0.017 | 0.124 | 0.261 | 0.191 | 0.205 | 0.277 | 0.216 | 0.121 | 0.111 | 0.135 | 0.135 | 0.199 |
| 2017 | 0.032 | 0.281 | 0.124 | 0.243 | 0.265 | 0.14 | 0.148 | 0.06 | 0.08 | 0.408 | 0.408 | 0.18 |
| 2018 | 0.017 | 0.141 | 0.132 | 0.127 | 0.408 | 0.21 | 0.37 | 0.204 | 0.145 | 0.836 | 0.836 | 0.227 |

Table 31.10. Sole in Division 7.e. Assessment summary.

| Year | Recruitment Age 2 [thousands] | TSB [tonnes] | SSB [tonnes] | Landings [tonnes] | Yield/SSB | Fbar |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| (Ages 3-9) |  |  |  |  |  |  |


| Year | Recruitment Age 2 [thousands] | TSB [tonnes] | SSB [tonnes] | Landings [tonnes] | Yield/SSB | Fbar <br> (Ages 3-9) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 3374 | 3861 | 2941 | 949 | 0.32 | 0.316 |
| 1998 | 4450 | 3983 | 2939 | 880 | 0.3 | 0.293 |
| 1999 | 3616 | 3997 | 2868 | 957 | 0.33 | 0.32 |
| 2000 | 6647 | 4375 | 2916 | 914 | 0.31 | 0.305 |
| 2001 | 5476 | 4594 | 2958 | 1069 | 0.36 | 0.342 |
| 2002 | 3864 | 4280 | 3091 | 1106 | 0.36 | 0.334 |
| 2003 | 5456 | 4520 | 3392 | 1078 | 0.32 | 0.25 |
| 2004 | 2896 | 4146 | 3211 | 1075 | 0.33 | 0.295 |
| 2005 | 4068 | 4121 | 3231 | 1039 | 0.32 | 0.323 |
| 2006 | 4698 | 3852 | 2863 | 1023 | 0.36 | 0.334 |
| 2007 | 4039 | 3975 | 2922 | 1015 | 0.35 | 0.344 |
| 2008 | 4426 | 4000 | 2832 | 908 | 0.32 | 0.311 |
| 2009 | 3943 | 4205 | 3206 | 701 | 0.22 | 0.206 |
| 2010 | 5038 | 4708 | 3664 | 698 | 0.19 | 0.196 |
| 2011 | 3725 | 5048 | 3825 | 801 | 0.21 | 0.192 |
| 2012 | 3597 | 4922 | 4053 | 872 | 0.22 | 0.212 |
| 2013 | 3260 | 4816 | 3956 | 883 | 0.22 | 0.216 |
| 2014 | 3215 | 5162 | 4277 | 885 | 0.21 | 0.196 |
| 2015 | 4354 | 5272 | 4307 | 774 | 0.18 | 0.176 |
| 2016 | 5735 | 6329 | 4843 | 913 | 0.19 | 0.199 |
| 2017 | 4061 | 6018 | 4502 | 1007 | 0.22 | 0.18 |
| 2018 | 5018 | 5934 | 4584 | 1075 | 0.23 | 0.227 |

Table 31.11. Sole in Division 7.e. Input data for the short-term forecast.

| Age | N2019 | N2020 | N2021 | M | Mat | PF | PM | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 3973 | 3973 | 3973 | 0.1 | 0.14 | 0 | 0 | 0.183 | 0.025 | 0.215 |
| 3 | 4464 | 3507 | 3483 | 0.1 | 0.45 | 0 | 0 | 0.245 | 0.205 | 0.275 |
| 4 | 2794 | 3291 | 2444 | 0.1 | 0.88 | 0 | 0 | 0.304 | 0.194 | 0.333 |
| 5 | 2765 | 2083 | 2326 | 0.1 | 0.98 | 0 | 0 | 0.359 | 0.21 | 0.386 |
| 6 | 1920 | 2028 | 1442 | 0.1 | 1 | 0 | 0 | 0.411 | 0.329 | 0.437 |
| 7 | 640 | 1250 | 1206 | 0.1 | 1 | 0 | 0 | 0.46 | 0.235 | 0.484 |
| 8 | 636 | 458 | 838 | 0.1 | 1 | 0 | 0 | 0.505 | 0.275 | 0.528 |
| 9 | 513 | 437 | 292 | 0.1 | 1 | 0 | 0 | 0.548 | 0.144 | 0.568 |
| 10 | 415 | 402 | 329 | 0.1 | 1 | 0 | 0 | 0.586 | 0.126 | 0.605 |
| 11 | 498 | 331 | 310 | 0.1 | 1 | 0 | 0 | 0.622 | 0.517 | 0.639 |
| 12 | 272 | 416 | 349 | 0.1 | 1 | 0 | 0 | 0.741 | 0.517 | 0.751 |

Table 31.12. Sole in Division 7.e. Single option output of the short-term forecast (targeting FMSY).

| Age | F | Catch.No | Yield | Stock.No | Biomass | SSNo | SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year $=$ 2019, $\mathrm{F} / \mathrm{F}_{2016-2018}=1.183, \mathrm{Fbar}=0.227$ |  |  |  |  |  |  |  |
| 2 | 0.025 | 93 | 20 | 3973 | 727 | 556 | 102 |
| 3 | 0.205 | 789 | 217 | 4464 | 1094 | 2009 | 492 |
| 4 | 0.194 | 469 | 156 | 2794 | 849 | 2459 | 747 |
| 5 | 0.21 | 499 | 193 | 2765 | 993 | 2709 | 974 |
| 6 | 0.329 | 514 | 224 | 1920 | 790 | 1920 | 790 |
| 7 | 0.235 | 128 | 62 | 640 | 295 | 640 | 295 |
| 8 | 0.275 | 146 | 77 | 636 | 321 | 636 | 321 |
| 9 | 0.144 | 66 | 37 | 513 | 281 | 513 | 281 |
| 10 | 0.126 | 47 | 28 | 415 | 243 | 415 | 243 |
| 11 | 0.517 | 192 | 123 | 498 | 310 | 498 | 310 |
| 12 | 0.517 | 105 | 79 | 272 | 202 | 272 | 202 |
| Total | NA | 3047 | 1216 | 18891 | 6104 | 12628 | 4756 |
| Year $=2020, F / F_{2016-2018}=1.183$, Fbar $=0.227$ |  |  |  |  |  |  |  |
| 2 | 0.032 | 118 | 25 | 3973 | 727 | 556 | 102 |
| 3 | 0.261 | 769 | 212 | 3507 | 859 | 1578 | 387 |
| 4 | 0.247 | 687 | 228 | 3291 | 999 | 2896 | 879 |
| 5 | 0.268 | 467 | 180 | 2083 | 749 | 2041 | 734 |
| 6 | 0.419 | 663 | 290 | 2028 | 834 | 2028 | 834 |
| 7 | 0.3 | 309 | 150 | 1250 | 576 | 1250 | 576 |
| 8 | 0.351 | 129 | 68 | 458 | 231 | 458 | 231 |
| 9 | 0.184 | 70 | 40 | 437 | 239 | 437 | 239 |
| 10 | 0.161 | 57 | 34 | 402 | 236 | 402 | 236 |
| 11 | 0.659 | 153 | 98 | 331 | 206 | 331 | 206 |
| 12 | 0.659 | 192 | 144 | 416 | 308 | 416 | 308 |
| Total | NA | 3613 | 1469 | 18176 | 5964 | 12393 | 4731 |
| Year $=2021, F / F_{2016-2018}=1.183$, Fbar $=0.227$ |  |  |  |  |  |  |  |
| 2 | 0.032 | 118 | 25 | 3973 | 727 | 556 | 102 |
| 3 | 0.261 | 764 | 210 | 3483 | 853 | 1568 | 384 |
| 4 | 0.247 | 510 | 170 | 2444 | 742 | 2151 | 653 |
| 5 | 0.268 | 521 | 201 | 2326 | 836 | 2280 | 819 |
| 6 | 0.419 | 472 | 206 | 1442 | 593 | 1442 | 593 |
| 7 | 0.3 | 298 | 144 | 1206 | 555 | 1206 | 555 |
| 8 | 0.351 | 237 | 125 | 838 | 424 | 838 | 424 |
| 9 | 0.184 | 47 | 27 | 292 | 160 | 292 | 160 |
| 10 | 0.161 | 47 | 28 | 329 | 193 | 329 | 193 |
| 11 | 0.659 | 143 | 91 | 310 | 192 | 310 | 192 |
| 12 | 0.659 | 161 | 121 | 349 | 259 | 349 | 259 |
| Total | NA | 3317 | 1349 | 16993 | 5535 | 11320 | 4334 |

Units are thousands (for numbers) and tonnes (for weights).

Table 31.13. Sole in Division 7.e. Year-class sources and contributions for the short-term forecast (in percent).

| cohort | Yield 2019 | Yield 2020 | SSB 2019 | SSB 2020 | SSB 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | 17.8 | 15.5 | 10.3 | 18.6 | 18.9 |
| 2017 | 1.6 | 14.4 | 2.1 | 8.2 | 15.1 |
| 2018 | 1.7 | 2.2 | 8.9 |  |  |
| 2019 |  | 2.3 |  |  |  |




Year class

|  | remaining YCs |
| :--- | :--- |
|  | YC 2016 |
|  | YC 2017 |
|  | YC 2018 |
|  | YC 2019 |

Table 31.14. Sole in Division 7.e. Annual catch scenarios. All weights are in tonnes.

| Basis | Total catch ${ }^{\wedge}$ (2020) | Wanted catch* (2020) | Unwanted catch* (2020) | $\begin{aligned} & F_{\text {wanted }} \\ & (2020) \end{aligned}$ | $\begin{aligned} & \text { SSB } \\ & \text { (2021) } \end{aligned}$ | \% SSB <br> change <br> ** | \% TAC change *** | \% Advice change <br> **** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |
| MSY approach: $\mathrm{F}_{\text {MSY }}$ | 1478 | 1469 | 9 | 0.29 | 4334 | -8.4 | 19.0 | 16.2 |
| $\mathrm{F}_{\text {MSY lower }}$ | 878 | 873 | 5 | 0.160 | 4915 | 3.9 | -29 | -31 |
| $\mathrm{F}_{\text {MSY upper }}$ | 1685 | 1675 | 10 | 0.34 | 4134 | -12.6 | 36 | 32 |
| Other options |  |  |  |  |  |  |  |  |
| $F=0$ | 0 | 0 | 0 | 0 | 5772 | 22 | -100 | -100 |
| $\mathrm{F}_{\mathrm{pa}}$ | 1603 | 1594 | 10 | 0.32 | 4213 | -11.0 | 29 | 26 |
| Flim | 2065 | 2053 | 13 | 0.44 | 3767 | -20 | 66 | 62 |
| SSB (2021) $=\mathrm{Bl}_{\text {lim }}$ | 3930 | 3906 | 24 | 1.20 | 2000 | -58 | 216 | 209 |
| $\operatorname{SSB}(2021)=B_{p a}=$ $\text { MSY } \mathrm{B}_{\text {trigger }}$ | 2972 | 2954 | 18 | 0.74 | 2900 | -39 | 139 | 134 |
| $F=F_{2018}$ | 1201 | 1194 | 7 | 0.23 | 4602 | -2.7 | -3.3 | -5.6 |
| Management plan: $\mathrm{F}=0.27$ with $15 \%$ TAC constraint | 1391 | 1383 | 9 | 0.27 | 4418 | -6.6 | 12.0 | 9.4 |

* "Wanted" and "unwanted" catch are used to describe fish that would be landed and discarded in the absence of the EU landing obligation, based on discard rate estimates for 2016-2018.
** SSB 2021 relative to SSB 2020 (4731 t).
*** Total catch in 2020 relative to TAC in 2019 (1242 t).
****Advice value 2020 relative to advice value 2019 (1272 t).
${ }^{\wedge}$ Total catch derived from the wanted catch and the unwanted catches ratio.

Table 31.15. Sole in Division 7.e. Annual catch scenarios (more options and more digits provided, sorted by fishing mortality in intermediate year). All weights are in tonnes.

| Basis | Total catch^ (2020) | Wanted catch* (2020) | Unwanted catch* (2020) | $F_{\text {wanted }}$ $(2020)$ | SSB <br> (2021) | $\begin{aligned} & \text { \% SSB } \\ & \text { change } * * \end{aligned}$ | \% TAC change *** | \% Advice change |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{F}=0$ | 0 | 0 | 0 | 0 | 5771.854 | 21.99769 | -100 | -100 |
| $\mathrm{F}_{\text {sa }} * 0.6$ | 759.8243 | 755.1817 | 4.642591 | 0.136488 | 5030.677 | 6.331692 | -38.8225 | -40.2654 |
| $\mathrm{F}_{\text {MSY lower }}$ | 878.4403 | 873.073 | 5.367345 | 0.16 | 4915.313 | 3.893278 | -29.2721 | -30.9402 |
| $\mathrm{F}=0.17$ | 927.8846 | 922.2151 | 5.669454 | 0.17 | 4867.254 | 2.877469 | -25.2911 | -27.0531 |
| $\mathrm{F}=0.18$ | 976.741 | 970.7731 | 5.967971 | 0.18 | 4819.783 | 1.874107 | -21.3574 | -23.2122 |
| Fsq0.8 | 986.3671 | 980.3403 | 6.026787 | 0.181984 | 4810.433 | 1.676461 | -20.5824 | -22.4554 |
| $\mathrm{F}=0.19$ | 1025.018 | 1018.755 | 6.262948 | 0.19 | 4772.894 | 0.88301 | -17.4704 | -19.4168 |
| TAC*0.85 | 1062.19 | 1055.7 | 6.490072 | 0.197782 | 4736.802 | 0.120146 | -14.4774 | -16.4945 |
| $\mathrm{F}=0.2$ | 1072.724 | 1066.17 | 6.554435 | 0.2 | 4726.576 | -0.096 | -13.6293 | -15.6664 |
| $\mathrm{F}=0.21$ | 1119.867 | 1113.024 | 6.842481 | 0.21 | 4680.821 | -1.06309 | -9.8336 | -11.9602 |
| $F=0.22$ | 1166.454 | 1159.327 | 7.127135 | 0.22 | 4635.623 | -2.01843 | -6.08258 | -8.29762 |
| $\mathrm{F}_{\text {sa }}$ | 1200.946 | 1193.608 | 7.337883 | 0.227481 | 4602.171 | -2.7255 | -3.30547 | -5.586 |
| $F=0.23$ | 1212.495 | 1205.086 | 7.408445 | 0.23 | 4590.972 | -2.96219 | -2.37564 | -4.6781 |
| TAC | 1249.635 | 1242 | 7.635378 | 0.238154 | 4554.965 | -3.72327 | 0.614765 | -1.75822 |
| $F=0.24$ | 1257.995 | 1250.309 | 7.686457 | 0.24 | 4546.862 | -3.89455 | 1.287854 | -1.10101 |
| $\mathrm{F}=0.25$ | 1302.964 | 1295.002 | 7.961219 | 0.25 | 4503.283 | -4.81565 | 4.908507 | 2.43425 |
| $F=0.26$ | 1347.408 | 1339.175 | 8.232775 | 0.26 | 4460.23 | -5.72566 | 8.486918 | 5.928264 |
| MP2 | 1391.334 | 1382.833 | 8.501169 | 0.27 | 4417.693 | -6.62473 | 12.02368 | 9.381608 |
| MP | 1391.334 | 1382.833 | 8.501169 | 0.27 | 4417.693 | -6.62473 | 12.02368 | 9.381608 |
| $F=0.27$ | 1391.334 | 1382.833 | 8.501169 | 0.27 | 4417.693 | -6.62473 | 12.02368 | 9.381608 |
| $\mathrm{F}=0.28$ | 1434.75 | 1425.984 | 8.766447 | 0.28 | 4375.667 | -7.51303 | 15.51936 | 12.79485 |
| TAC*1.15 | 1437.081 | 1428.3 | 8.780685 | 0.28054 | 4373.412 | -7.56069 | 15.70698 | 12.97804 |
| $\mathrm{F}_{\text {MSY }}$ | 1477.664 | 1468.635 | 9.028652 | 0.29 | 4334.144 | -8.39069 | 18.97454 | 16.16854 |
| $\mathrm{F}=0.3$ | 1520.081 | 1510.793 | 9.287825 | 0.3 | 4293.116 | -9.25788 | 22.38978 | 19.50323 |
| $\mathrm{F}=0.31$ | 1562.009 | 1552.465 | 9.544009 | 0.31 | 4252.577 | -10.1147 | 25.76563 | 22.79946 |
| $\mathrm{F}_{\mathrm{pa}}$ | 1603.455 | 1593.658 | 9.797245 | 0.32 | 4212.521 | -10.9614 | 29.10264 | 26.05777 |
| $\mathrm{F}=0.32$ | 1603.455 | 1593.658 | 9.797245 | 0.32 | 4212.521 | -10.9614 | 29.10264 | 26.05777 |
| $\mathrm{F}=0.33$ | 1644.425 | 1634.377 | 10.04757 | 0.33 | 4172.939 | -11.798 | 32.40133 | 29.27866 |
| $\mathrm{F}_{\text {MSY upper }}$ | 1684.925 | 1674.63 | 10.29504 | 0.34 | 4133.826 | -12.6247 | 35.66224 | 32.46266 |
| $\mathrm{F}_{\text {lim }}$ | 2065.482 | 2052.862 | 12.62027 | 0.44 | 3767.091 | -20.3763 | 66.3029 | 62.38067 |
| $\mathrm{B}_{\mathrm{pa}}$ | 2972.39 | 2954.229 | 18.16156 | 0.735613 | 2900 | -38.7037 | 139.3229 | 133.6785 |
| $\mathrm{B}_{\text {trigger }}$ | 2972.39 | 2954.229 | 18.16156 | 0.735613 | 2900 | -38.7037 | 139.3229 | 133.6785 |
| Blim | 3929.8 | 3905.789 | 24.01141 | 1.196931 | 2000 | -57.7267 | 216.409 | 208.9465 |

* "Wanted" and "unwanted" catch are used to describe fish that would be landed and discarded in the absence of the EU landing obligation, based on discard rate estimates for 2016-2018.
** SSB 2021 relative to SSB 2020.
*** Total catch in 2020 relative to TAC in 2019 (1242 t).
****Advice value 2020 relative to advice value 2019 (1272 t).
${ }^{\wedge}$ Total catch derived from the wanted catch and the unwanted catches ratio.


Figure 31.1. Sole in Division 7.e. Reported landings and discards by country.


Figure 31.2. Sole in Division 7.e. International landings by fleet and year.


Figure 31.3. Sole in Division 7.e. Discard rates for discards reported in InterCatch.


Figure 31.4. Sole in Division 7.e. Annual reported discard rates in InterCatch by fleet and country.


Figure 31.5. Sole in Division 7.e. International landings numbers-at-age (last 16 years).


Figure 31.6. Sole in Division 7.e. Catch (landings) and stock weights-at-age.


Figure 31.7. Sole in Division 7.e. Generation of stock and catch weights from landings weights-at-age.


Figure 31.8. Sole in Division 7.e. Landings age distributions from InterCatch provided by France. Numbers are raised to fleet level.


Figure 31.9. Sole in Division 7.e. Landings age distributions submitted to InterCatch. Numbers are raised to fleet level.


Figure 31.10. Sole in Division 7.e. Length distributions from samples for France. Numbers are raised to fleet level.


Figure 31.11. Sole in Division 7.e. Length distributions from samples for UK (England). Numbers are raised to fleet level.


Figure 31.12. Sole in Division 7.e. Means standardised lpue and effort for the UK commercial fleets.


Figure 31.13. Sole in Division 7.e. Means standardised lpue/cpue by year class. Note, the cohorts differ on the $x$-axes due to the differences in the length and age ranges of the tuning series.


Figure 31.14. Sole in Division 7.e. Means standardised lpue/cpue by year. Note, the lines differ on the x-axes due to the differences in the length and age ranges of the tuning series.


Figure 31.15. Sole in Division 7.e. Survey indices (raw values) for all commercial and scientific surveys. The plots on the left show the index values at age, on the right are the values aggregated over all ages.


Figure 31.16. Sole in Division 7.e. Internal consistencies in the scientific surveys. Shown is the correlation between numbers-at-age and the numbers of the same cohort one year later, including Pearson correlation coefficient $\rho$ and the $p$-value.


Figure 31.17. Sole in Division 7.e. Internal consistencies in the commercial surveys. Shown is the correlation between numbers-at-age and the numbers of the same cohort one year later, including Pearson correlation coefficient $\rho$ and the $p$-value.


Figure 31.18. Sole in Division 7.e. Results of the final XSA run. ICES estimated catches, recruitment, fishing mortality, and spawning-stock biomass from the summary stock assessment. Assumed recruitment values are not shaded. Discard estimates are only available from 2012 onwards.





| FSP-UK |  |
| :---: | :---: |
|  | 00000000000000 |
|  | 00000000000000 |
|  | $000000000000 \cdot 0$ |
|  | $000000 \cdot 0000000$ |
|  | $00000000000 \cdot \circ 00$ |
|  | $0 C 000000000000$ |
|  | 00000000000 |
|  | $000000000 \cdot 0000$ |
|  | 0000000160000 |
|  | cemboon 000 |
| 1990 | 2010 |


log residuals

- 0.01
- 0.10

○ 0.50
○ 1.00
$\bigcirc 2.00$

O negative

- positive
age

11
10

- 8

7

- 6
$-5$
- 3
- 2

Figure 31.19. Sole in Division 7.e. XSA fleet log catchability residuals. Note that the application of time-series weighting set as a tricubic taper with a range of 15 years excludes log catchability residuals for the UK-COT fleet prior to 2004.


Figure 31.20. Sole in Division 7.e. Comparison of the current XSA assessment with the final assessment runs from the last two years.


Figure 31.21. Sole in Division 7.e. Scaled weights for the current XSA assessment and the previous XSA assessment conducted at last year's WGCSE.


Figure 31.22. Sole in Division 7.e. Five-year retrospective of stock status and fishing mortality estimates.


Figure 31.23. Sole in Division 7.e. Options for the intermediate year in the short-term forecast.


Figure 31.24. Sole in Division 7.e. Output for the short-term forecast under the MSY approach.


Figure 31.25. Sole in Division 7.e. Output of the short-term forecast of the MSY approach, including Fmsy ranges.


Figure 31.26. Sole in Division 7.e. Output of the short-term forecast of the MSY approach, including Fmsy ranges zoomed into the most recent years.

## 34 Sole in divisions 7.f and 7.g (Bristol Channel, Celtic Sea)

The assessment of sole in division 7.f and 7.g presented at WGCSE 2019 is the first assessment after the Inter-benchmark (IBPBrisol, ICES 2019) in February 2019. The Stock Annex was updated with respect to the outcomes of the Inter-benchmark. Changes include the use one adjusted commercial tuning series and a new selection of ages for the tuning fleets.

## Type of assessment in 2019

This assessment is an update assessment.

## ICES advice applicable to 2018

In the advice for 2018, the stock status was presented as follows:


ICES advises that when the MSY approach is applied, catches in 2017 should be no more than 931 tonnes.

## ICES advice applicable to 2019

In the advice for 2019, the stock status was presented as follows:


ICES advises that when the MSY approach is applied, catches in 2018 should be no more than 864 tonnes.

### 34.1 General

## Stock description and management units



A TAC is in place for ICES divisions 7.f and 7.g. These divisions do correspond to the stock area. The basis for the stock assessment area 7.f and 7.g is described in detail in the Stock Annex.

## Management applicable to 2018 and 2019

The sole fisheries in the Celtic Sea are managed by TAC and technical measures. The agreed TACs in 2018 and 2019 are presented in the text tables below. Technical measures in force for this stock are minimum mesh sizes and minimum conservation reference sizes (MCRS, 25 cm for Belgian vessels from March 11th 2017 onwards, except vessels with engine power <221 kW and/or volume $<70 \mathrm{GT}$ ). National regulations also restricted areas for certain types of vessels.

2018 TAC

|  |  |  |  |
| :--- | :--- | :--- | :--- |
| Species: | Common sole <br> Solea solea | Zone: | 7 f and 7 g <br> (SOL $/ 7 \mathrm{FG})$. |
| Belgium | 574 |  |  |
| France | 58 |  |  |
| Ireland | 29 |  |  |
| United Kingdom | 259 |  |  |
| Union | 920 | Analytical TAC |  |
| TAC | 920 |  |  |

2019 TAC

| Species:Common sole <br> Solea solea | Zone:7f and 7 g <br> (SOL/7FG.) |  |
| :--- | ---: | :--- | :--- |
| Belgium | 525 |  |
| France | 53 |  |
| Ireland | 26 |  |
| United Kingdom | 237 |  |
| Union | 841 | Analytical TAC |
| TAC | 841 |  |

Three rectangles in the Celtic Sea (30E4, 31E4 and 32E3, referred to as the "Trevose Box") were closed during the first quarter of 2005, and in February-March each year from 2006 onwards. A derogation has permitted beam trawlers to fish there in March 2005. The effects of this closure
have been discussed in previous WGSSDS meetings and ACFM 2007, and evaluated at WKCELT 2014.

## Fishery in 2018

The Expert Group estimated the total international landings at 850 t in 2018 (Table 36.1), of which Belgium landed $71 \%$ ( 607 t ), UK $20 \%$ ( 171 t ), France 5\% (44 t), and 3\% by Ireland ( 28 t ). This landing figure corresponds to an international uptake of $92.4 \%$ of the agreed TAC in 2018 ( 920 t ).

In $2018,90 \%$ of the landings were taken by beam trawls, $9.3 \%$ by otter trawls, $0.59 \%$ by other gears.

Early in the time-series officially reported landings included divisions 7.g-k for some countries and their total was higher than the WG estimate. Since 1999, official landings correspond to divisions 7.f and 7.g, and the total is lower than the working group estimate. During the period 2002-2005 the difference between the two estimates was substantial. This was mainly due to area misreporting, which was taken into account in the working group estimates (WKCELT ICES, 2014). In the recent years, the estimates are more similar.

### 34.2 Data

## Landings

Annual length compositions for 2018 are given by fleet in Table 36.2. Length distributions of the total Belgian and UK(E\&W) landings for the last nineteen years are plotted in Figure 36.1. Belgian vessels land a greater proportion of small fish compared to the UK(England \& Wales).

Belgium, Ireland, France and UK have provided data this year under the ICES InterCatch format on a métier basis. Quarterly/yearly data for 2018 were available for landing numbers and weight-at-age, for most of the Belgian, Irish and UK fleets. These comprise $90 \%$ of the international landings. Allocation has been made as follows: seven groups of métiers with age distributions were set up: e.g. OTB_DEF_70-99, OTB_DEF_100-119, OTB_CRU_70-99, OTT_CRU_100-119, OTT_DEF_100-119, GNS_DEF_all and a group of all available métiers with age distributions (Overall). The OTB_DEF_70-99 (1.9\% of overall landings), OTB_DEF_100-119 (2.0\% of overall landings), OTB_CRU_70-99 (4.9\% of overall landings), OTT_CRU_100-119 ( $0.4 \%$ of overall landings), OTT_DEF_100-119 (2.0\% of overall landings), and GNS_DEF_all ( $<1 \%$ of overall landings) métiers without age distributions were allocated with the group OTB_DEF_70-99, OTB_DEF_100-119, OTB_CRU_70-99, OTT_CRU_100-119, OTT_DEF_100-119, GNS_DEF_all respectively. The rest of the métiers without age distributions ( $2.0 \%$ of overall landings) were allocated to the group Overall.

For the period 2008-2018, the original total international catch weights-at-age were used. The stock weights were obtained using the Rivard weight calculator (http://nft.nefsc.noaa.gov./), that conducts a cohort interpolation of the catch weights.

Catch numbers-at-age are given in Table 36.3, and weights-at-age in the catch and the stock are given in Tables 36.4-36.5. Age compositions over the last eighteen years are plotted in Figure 36.2. The standardised catch proportion-at-age is presented in Figure 36.3.

The low catch numbers for age 1 in 2014 and 2016 were excluded from the catch numbers-at-age matrix to be consistent with previous years for which age 1 catch numbers are set to zero.

## Discards

The available discard data indicate that discarding of sole has increased in 2017 and 2018. The length distributions for 2018 of retained and discarded catches of sole are presented in Figure
36.4a from the Belgium beam trawl fleet, in Figure 36.4b from the UK beam trawlers and in Figure 36.4 c from the Irish beam and otter trawlers. It should be noted that the Irish otter trawl landings only amount to about $2.2 \%$ of the total international landings.

As an attempt, estimating an overall discard rate for the stock, individual discard estimates for 2016-2018 from the main métiers and countries were averaged to arrive at an overall discard rate by year (Table 36.6). The percent of the métiers with discard information covering the total international landings is, $87 \%, 90 \%$ and $90 \%$ for 2016, 2017 and 2018 respectively.

The average discard rate for 2016-2018 is $8.9 \%$, while the discard rate for 2018 is $14.8 \%$. This steep increase is due to Belgian and Irish beam trawlers. The Belgian beam trawl fleet mainly discarded fish of 22 and 23 cm (Figure 36.4a). According to the Belgian age-length samples, these fish were mainly age 2 in 2018. These discarded fish thus belong to the very strong 2016 year class. Given the high recruitment of this year class, it makes sense that there is an increase in discarding. The Irish beam trawl fleet discards sole over the whole length range (Figure 36.4c) due to quota restrictions.

## Biological

Natural mortality was assumed to be 0.1 for all ages and years. The maturity ogive is based on samples taken during the UK(E\&W) beam trawl survey of March 1993 and 1994 and is applied to all years of the assessment.

The proportion of M and F before spawning was set to zero.

## Surveys

Standardised abundance indices for the UK beam trawl survey (UK(E\&W)-BTS-Q3)) are shown in Table 36.7 and Figure 36.5. Abundance-at-age 0 is highly variable and not used further on. The UK-survey appears to track the stronger year classes reasonably well. The internal consistency plot indicates also a reasonable fit for most of the ages (Figure 36.6).

## Commercial Ipue

Available estimates of effort and lpue are presented in Tables 36.8-36.9 and Figure 36.7.
Commercial lpue and effort data were available for Belgian beam trawlers, UK (E\&W) beam and otter trawlers and Irish seiners, otter and beam trawlers. It should be noted that the most recent lpue values of the UK (E\&W) beam trawlers (2013-2018) and the UK (E\&W) otter trawlers (20142018) are not available due to reporting issues. In 2013, the UK administration switched to the EU electronic logbook system. Therefore, a lot of the reported effort is missing and the 2013 value cannot be used as an absolute number. Details of the 2013 UK beam trawl were unavailable due to reduced numbers of trips reporting this gear specific effort information via the newly introduced e-logbook system. The otter trawl fleet effort reporting was unaffected by this as these vessels were not reporting their landings via this method in 2013. However, from 2014 onwards both the UK beam trawl and otter trawl effort values are unavailable because of the reporting issues. Therefore, the UK(E\&W)-CBT tuning indices for the six most recent years were excluded in this year's assessment.

Belgian beam trawl effort was at highest levels in 2003-2005. During these years effort shifted from the Eastern English Channel (Division 7.d) to the Celtic Sea (divisions 7.fg) because of days-at-sea limitations in the former area. In 2006, these restrictions had been lifted and effort decreased substantially to about half of the values observed in the early 2000s. The sharp effort reduction in 2008 may be a combined result of the unrestricted effort regime in Division 7.d and the high fuel prices. The increase in 2012-2013 is due to the good opportunities of sole catches in
the Celtic sea taken by the mobile Belgian fleet. Afterwards, effort decreased again to the lower level recorded in 2010.

The effort from the UK(E\&W) beam trawl fleet has declined sharply since the early 2000s to a record low in 2009, and stayed at that level since. Inspection of an alternate effort indicator (days fished) suggests that the beam trawl effort in 2013-2018 remains low.

The effort from the UK (E\&W) otter trawlers has shown a gradually declining trend over time and the alternate effort indicator (days fished) suggests that this declining trend continues in the period 2013-2018.

Lpue of the Belgian beam trawlers peaked in 2002. After a sharp decline to its record low in 2004, lpue has been increasing gradually, levelling off in 2010-2013 at around $15 \mathrm{~kg} / \mathrm{hour}$. In 20142015, lpue increased to around $19-20 \mathrm{~kg} / \mathrm{hour}$. In 2016 and 2017, a decrease to $15.63-15.08$ $\mathrm{kg} /$ hour was recorded. In 2018, lpue increased to $18 \mathrm{~kg} / \mathrm{hour}$.

Lpue of the UK beam trawlers was stable in the 1990s and 2000s, but at lower levels compared to the period before. In 2007, lpue increased considerably and gave a similar value for 2008. In 2009, there was a decrease to a level just above the mean of the time-series, followed by similar values for 2010, 2011 and 2012. Inspection of the alternate lpue indicator ( $\mathrm{kg} /$ days fished) suggests that the beam trawl lpue in 2014-2018 substantially increased.

In the period 1987-2010, the lpue of the UK otter trawlers remains stable at a lower level. In 20112013, the lpue increased to the higher level that was recorded in the beginning of the eighties. The alternate lpue indicator (kg/days fished) suggests a similar level in 2018.

Irish effort and lpue data are also presented. The main target species in the Irish fisheries are megrim, anglerfish, etc. The vessels usually operate on fishing grounds in the Western Celtic Sea with lower sole densities and therefore the lpue values are low.

## Tuning series

During the 2019 IBP, the Belgian commercial beam trawl tuning fleet (BE-CBT) was substantially revised. Prior to the IBP, the BE-CBT tuning series consisted of two parts, which were included separately in the assessment: one with the original data from 1971 up to 1996 and one series with data from 1997 up to 2017. For the latter, the effort was corrected for engine power, based on a study carried out by IMARES and CEFAS in the mid-1990s (applicable to sole and plaice effort in the beam trawls fisheries). Currently, this method is outdated and during the IBP, a more realistic conversion factor for engine power was investigated to convert nominal fishing effort to effective effort.

To calculate a new Belgian beam trawl tuning series (BE_CBT3), the sales notes and logbooks of the Belgian beam trawl fleet were used to calculate the sole landing rates in the Celtic Sea. To account for misreporting, only the data from vessels with HP $>221 \mathrm{Kw}$ and only those fishing trips in which fishing activity was limited to the Celtic Sea were retained for statistical analysis. A GLMM model with all explanatory variables (year, month, ICES statistical rectangle and loglinear effect of a vessel's engine power), a random vessel effect, a variable dispersion parameter governed by monthly, and spatial effects, a logarithmic link function between the linear predictors and response variable, and a Gamma distributed error term showed the best model fit.

The exponent of the estimated coefficients of the year effect were used as landing rate for the tuning series. To convert this landing rate per year to the annual sole age compositions (age 1 to 15 (plus group)) of the Belgian beam trawlers (TBB_DEF_70-99) active in the ICES divisions 27.7.f and 27.7.g, it was standardised by the total weight landed by the pure trips of the large fleet segment per year.

During the IBP, it was decided to include the new Belgian tuning series (BE_CBT3) from 2006 up until the last data year with ages 2-9. The old Belgian CBT from 1971-1996 was trimmed to ages $3-9$. The BE_CBT2 series running from 1997 up until the last data year was excluded. Finally, the UK(E\&W)-CBT from 1991-2012 was also trimmed to ages 3-8. Due to effort reporting issues, the 2013-2018 UK-CBT indices were not available and could not be used in the assessment. Settings for the UK BTS Q3 survey remained unchanged. More information is provided in the stock annex and the 2019 IBP report (ICES, 2019b).
The internal consistency plots for the main two commercial lpue series, used in the assessment (UK(E\&W)-CBT(1991-2012), BEL-CBT(1971-1996) and BEL-CBT3 (2006-2018)), show high consistencies for the entire age range (Figures 36.8-36.9).

## Other relevant data

Reports from UK industry suggest that the main issues affecting the fishery in 7.f and 7.g were displacement of effort due to the rectangle closures and the restrictions on the use of 80 mm mesh west of $7^{\circ} \mathrm{W}$ (Trebilcock and Rozarieux, 2009).

No additional information was received from the Belgian, French and Irish industries.

### 34.3 Stock assessment

The method used to assess Celtic Sea sole is XSA, using one survey and two commercial tuningseries (Table 36.10). The Belgian commercial beam trawl tuning fleet is now split into two parts (period 1971-1996 and 2006-2018). Table 36.10 also includes tuning indices of the Irish ground fish survey (IGFS-IBTS_Q4), the commercial UK otter trawl fleet (UK(E\&W)-COT) and the Belgian commercial beam trawl tuning fleet for 1997-2005 which are not used in this assessment.

## Data screening

As mentioned in Section 36.2, the 2013-2018 data from the UK(E\&W) commercial tuning series were excluded from the assessment.

There has been a substantial change in perception of the stock due to an Inter-benchmark in which a new Belgian commercial tuning index was constructed focusing on the landings and effort data of trips entirely within divisions 7.f and 7.g. The selected ages of the different tuning series were revised. These changes resulted in a reduced retrospective bias. These changes cause a substantial upward revision of the stock size and a substantial downward revision of the F. In addition, the recruitment of 2015 and 2017 are among the highest in the time-series.

## Final update assessment

The final settings used in this year's assessment are as detailed below:

|  | 2019 assessment |  |  |
| :---: | :---: | :---: | :---: |
| Fleets: | Years | Ages | \| $\alpha-\beta$ |
| BEL-CBT commercial | 1971-1996 | 3-9 | 0-1 |
| BEL-CBT3 commercial | 2006-2018 | 2-9 | 0-1 |
| UK-CBT commercial | 1991-2012 | 3-8 | 0-1 |
| UK(E\&W)-BTS-Q3 survey | 1988-2018 | 1-5 | 0.75-0.85 |
| -First data year | 1971 |  |  |
| -Last data year | assessment year-1 |  |  |
| -First age | 1 |  |  |
| -Last age | 10+ |  |  |
| Time-series weights | None |  |  |
| -Model | Mean q model all ages |  |  |
| -Q plateau set at age | 7 |  |  |
| -Survivors estimates shrunk towards mean F | 5 years / 5 ages |  |  |
| -s.e. of the means | 1.5 |  |  |
| -Min s.e. for pop. Estimates | 0.3 |  |  |
| -Prior weighting | None |  |  |
| Fbar | Ages 4-8 |  |  |

The catchability residuals for the final XSA are shown in Figure 36.10 and the XSA tuning diagnostics are given in Table 36.11. There may be some indications of a trend in the UK beam trawl fleet (UK(E\&W)-CBT) with predominantly positive residuals around 2007.

F shrinkage gets low weights (maximum 4,5\%) in determining the survivor estimates for all ages. The weighting of the UK(E\&W)-BTS-Q3 decreases for the older ages as only the tuning indices for the younger ages are used in the assessment (age range: $1-5$ ). The estimates for the recruiting age are solely defined by the survey. The commercial fleet BE-CBT3 on the other hand are given more weight (Figure 36.11) for the older ages.

The UK(E\&W)-BTS-Q3 is rather consistent in predicting year-class strengths at different ages (Figure 36.5), where the UK and Belgian (new) commercial tuning series have a higher variability in estimating the year-class strength at different ages.

Retrospective patterns for the final run are shown in Figure 36.12. There appears to be a small retrospective bias in estimating F and SSB in the most recent years, at which F was underestimated and SSB was overestimated.

A Mohn's rho analysis was conducted based on the XSA stock assessment results, i.e. the last data year (2018) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (ages 4-8) | recruitment |
| :--- | :--- | :---: | :--- |
| Mohn's rho value | 0.075 | -0.038 | 0.045 |

The Mohn's rho values for this assessment are low and well below the threshold of $20 \%$ imposed by ICES for 2019 assessments, i.e. the current assessment indicates a high consistency.

The final XSA output is given in Table 36.12 (fishing mortalities) and Table 36.13 (stock numbers). A summary of the XSA results is given in Table 36.14 and trends in yield, fishing mortality, recruitment and spawning stock biomass are shown in Figure 36.13.

## Comparison with previous assessment

A comparison of the estimates of this year's assessment with last year's is given in Figure 36.14. With Inter-benchmark in 2019 and the addition of the 2018 data, there is a substantial upward revision of the SSB and a strong downward revision of the F.

## State of the stock

Trends in landings, SSB, $\mathrm{F}(4-8)$ and recruitment are presented in Table 36.14 and Figure 36.13.
During the eighties fishing mortality increased for this stock. In the following decades fishing mortality fluctuated above FmSY. Since 2006, fishing mortality decreased and fluctuated between FMSY and $F_{\text {pa. }}$. In 2012 fishing mortality began to increase again, then decreased and is below FMSY since 2017.

Recruitment has fluctuated around 5 million recruits with occasional strong year classes. The 1998 year class is estimated to be the strongest in the time-series and the 2007 year class is also one of the stronger year classes. The 2009 year class is by the lowest in the time-series. The 2014 and 2016 year classes are estimated to be among the highest of the time-series.

SSB has declined almost continuously from the highest value of 7385 t in 1971 to the lowest observed in the time-series in 1998 ( 1592 t ). The exceptional year class of 1998 has increased SSB to above the long-term average. Spawning-stock biomass (SSB) has been above MSY Btrigger since 2001. The SSB increased during the last years as a result of the decreasing fishing mortality and continuous good recruitment.

### 34.4 Short term projections

The long-term GM71-16 recruitment (4975 thousand fish) was assumed for recruitment in 2019 and subsequent years.

Population numbers at the start of 2019, estimated for ages 2 and older, were taken from the XSA output.

Fishing mortality was set as the mean over the last three years scaled to 2018. Weights-at-age in the catch and in the stock are averages for the years 2016-2018. Input to the short-term predictions, the sensitivity analysis and the FMSY analysis are shown in Table 36.15. Results are presented in Table 36.16 (management options) and Table 36.17 (detailed output).

The working group decided to use a TAC constraint for the intermediate year (2019) as recent landings have been close to the TAC or only limited overshot. Moreover, status quo fishing mortality gives significantly higher landings ( 1187 t ) in the intermediate year than the agreed TAC (841 t).

Assuming a TAC constraint for 2019 of 841 t ( 766 t landings), implies a fishing mortality in 2019 of 0.158 . The assumed landings using a status quo fishing mortality in 2020 is 1252 t . This results in a SSB of 5439 t in 2020 and 5523 t in 2021.

Assuming a TAC constraint for 2019 and a status quo F in 2020, the proportional contributions of recent year classes to the predicted landings and SSB are given in Figure 36.15. The assumed GM recruitment accounts for about $2.3 \%$ of the landings in 2020 and about $11.3 \%$ of the 2021 SSB.

There are no known specific environmental drivers known for this stock.

### 34.5 MSY explorations

Investigations for possible Fmsy candidates for this stock were done at WGCSE 2010. ACOM adopted an FMSY value of 0.31, based on stochastic simulations using a "Ricker" model (PLOTMSY program). Btrigger was set to the $B_{p a}$ value of 2200 t .

Exploratory analysis investigating possible revisions of MSY estimates were conducted at WGCSE 2014 with a recent version of PLotMSY (Cefas, 2014). The simulations indicated that there is no reason for using a particular weighting for any of the stock-recruitment relationships. The resulting FMSY values were in line with the FMSY of 0.31 used at that moment for this stock.

In response to the EC long-term management plans for western EU waters (ICES subareas 5 to 10), ICES WKMSYREF4 (October 2015, Brest (France)) used long-term stochastic simulations (Eqsim) to estimate FMSY and appropriate ranges. The methodology used for stocks with agebased assessments follows the approaches developed in ICES WKMSYREF2 (ICES, 2014b) and WKMSYREF3 (ICES, 2014c) and is documented in the report of WKMSYREF4 (ICES, 2016). Estimates of reference points $\mathrm{Blim}_{\mathrm{lim}} \mathrm{B}_{\mathrm{pa}}, \mathrm{F}_{\text {lim }}$ and $\mathrm{F}_{\mathrm{pa}}$ were provided, and the $\mathrm{F}_{\mathrm{MSY}}$ ranges [ $\mathrm{F}_{\text {lower, }} \mathrm{F}_{\text {upper }}$ ] deliver no more than $5 \%$ reduction in long-term yield compared with MSY.

The full available time-series of sole 7.f and 7.g recruitment was used to fit stock-recruitment models. The simulations indicated that there is no reason for using a particular weighting for any of the stock-recruitment relationships. The workshop decided to use a more conservative approach and to base the analysis on a segmented regression only with a breakpoint set at Blim of 1700 t . Blim was chosen as the lowest value of the SSB time-series (Bloss). The revised MSY reference points are more restrictive ( $\mathrm{F}_{\mathrm{MSY}}=0.27$ instead of 0.31 and MSY $B_{\text {trigger }}=2400 \mathrm{t}$ instead of 2200 t ) and demand a larger reduction in F to achieve the MSY objectives as foreseen in the basic regulation.

The Inter-benchmark (IBPBrisol, ICES, 2019) for sole in divisions 27.7 f and 27.7 g updated MSY and PA reference points according to ICES guidelines. The results of the IBP were presented during the 2019 WGCSE assessment working group. However, the proposed reference points were rejected by the WG and re-estimated during the WG. The Fmsy value was perceived too high compared to the value prior to the inter-benchmark. The main reason was that a decreasing trend in recruitment when SSB is high (Ricker model), was not considered realistic for this stock. During the IBP, the FmSY value was estimated based on both the Ricker and Segmented Regression models. The WGCSE 2019 suggested to only consider the segmented regression model.

A working document describes the new calculation of the reference points taking into account the remarks from the WGCSE 2019 (ICES. 2019. WGCSE. ICES Scientific Reports. 1:29. xx pp.).

### 34.6 Biological reference points

## Precautionary approach reference points

The Working Group's current approach to reference points is outlined in Section 7.13.5. Current biological reference points calculated during the WGCSE 2019 are given in the text table below:

| Framework | Reference <br> point | Value | Technical basis |
| :--- | :--- | :--- | :--- |
| MSY approach | MSY $B_{\text {trigger }}$ | 2228 tonnes | $\mathrm{B}_{\text {pa }}$ |
| Precautionary <br> approach | $\mathrm{B}_{\text {lim }}$ | 0.297 | EQsim analysis based on the recruitment period 1971-2017 |

* EU multiannual plan (MAP) for the Western Waters (EU, 2019).


### 34.7 Management plans

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to demersal stocks including sole in ICES divisions 7.f and 7.g.

### 34.8 Uncertainties and bias in assessment and forecast

## Sampling

The major fleets fishing for 7.f and 7.g sole are sampled (approximately $90 \%$ of the total landings). Sampling is considered to be at a reasonable level.

## Discards

Discard estimates used to be low, but are increasing. Discards are not included in the assessment.

## Surveys

The UK(E\&W)-BTS-Q3 survey, which is solely responsible for the recruiting estimates, has been able to track year-class strength at-ages greater than 0 rather well in the past. However, the estimates of strong year classes have sometimes been revised downward in previous assessments and may cause bias in the forecast.

## Consistency

The assessment provided by the Expert Group shows a substantial upward revision of the SSB and a substantial downward revision of the $F$. The new assessment shows a low retrospective bias and is thus fairly consistent compared to the assessment of the last year's working group.

## Misreporting

Area misreporting is known to have been considerable over the period 2002-2005. This was due to a combination of the good 1998 year class still being an important part of the catch composition and more restrictive TACs. The area misreporting has been corrected for the years 2002-2006 (WGSSDS 2007). At the WKCELT 2014, analysis revealed that there was additional misreporting taking place in 2002-2003 and 2004 which was not accounted for in the first correction done at WGSSDS in 2007. Since 2007 the area misreporting that could be estimated was negligible.

### 34.9 Recommendation for next Benchmark

Sole in 7.f and 7.g has been benchmarked in February 2014. A new benchmark is scheduled for 2020 (WKFlatC). The issues are listed below.

Tuning series

| Problem / Aim | Work needed / Work needed / possible direction of solution | Data needed to be able to do this: are these available / where should these come from? |
| :---: | :---: | :---: |
| Commercial UK(E\&W)-CBT fleet <br> The UK beam trawl tuning-series is in the current assessment used up to 2012, because of effort reporting issues. A new tuning series was provided with effort in days instead of hours up to 2015. The inclusion of this new tuning series results in a significant upward revision of F and downward revision of SSB from the late 90's up until now, compared to the original tuning series. | *Need to review the new UKCBT tuning series with effort in days | *UK-CBT tuning series calculations |
| UK-BTS-Q3 survey <br> The UK-BTS-Q3 survey is the only survey used in the current assessment and is solely providing information on the recruiting age (age 1) | *Investigate if additional survey information (e.g. UKQ1SWBeam, started in 2006) is available and can be incorporated in the assessment. <br> *Additional survey data can confirm the info provided by the UK-BTS-Q3 survey. | *UK-Q1SWBeam tuning series <br> *other available survey data |

Fisheries \& ecosystem issues and data:

| Trends in mean weights | ${ }^{*}$ What drives this change? | information on the evolution in <br> the Celtic Sea ecosystem |
| :--- | :--- | :--- |
| The mean weights have dropped over time (2000- <br> 2010) and recently increased again. | *Is it driven by an ecosystem <br> change? | *Is there a similar trend in the <br> weights from other stocks? |


| Alternative assessment models to XSA. | *Explore the use of A4A, ASAP | *Standard assessment inputs |
| :--- | :--- | :--- |
| The current assessment has a developing |  |  |
| and SAM as alternatives to XSA |  |  |
| retrospective pattern that could create issues in the |  |  |
| forecast. |  |  |
| It would be preferable to use a statistical method <br> and propagated the main uncertainties into the <br> forecasts properly. |  |  |

### 34.10 Management considerations

Following the recent strong year classes, SSB is now at its highest level since 1975.
The Celtic Sea is an area without days-at-sea limitations for demersal fisheries. In this context and given that many demersal vessels are very mobile, changes in effort measures in areas other than the Celtic Sea, can influence the effort regime in the Celtic Sea (cfr. increased effort in Celtic Sea for Belgian beamers during 2004-2005 when days-at-sea limitations were in place for the Eastern English Channel).

### 34.11 Ecosystem considerations

Sole and plaice are predominantly caught by beam trawl fisheries. Beam trawling is known to have an impact on the benthic communities, although less so on soft substrates and in areas which have been historically exploited by this fishing method. Benthic drop-out panels have been shown to release around $75 \%$ of benthic invertebrates from the catches. Information from the UK industry (Trebilcock and Rozarieux, 2009) suggests that uptake in 2008 was minimal.

### 34.12 References

EU. 2019. REGULATION (EU) 2019/472 OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. 17pp. https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX\%3A32019R0472.

ICES. 2009. Report of the Benchmark and Data Compilation Workshop for Flatfish (WKFLAT 2009), 6-13 February 2009, Copenhagen, Denmark. ICES CM 2009/ACOM:31. 192 pp.
ICES. 2014. Report of the Benchmark WKCELT, 3-7 February 2014, Copenhagen, Denmark. ICES CM 2014/ACOM:42. 194 pp.

ICES. 2014b. Report of the Workshop to consider reference points for all stocks (WKMSYREF2), 8-10 January 2014, ICES Headquarters, Copenhagen, Denmark. ICES CM 2014/ACOM:47. 91 pp.

ICES. 2014c. Report of the Joint ICES-MYFISH Workshop to consider the basis for Fmsy ranges for all stocks (WKMSYREF3), 17-21 November 2014, Charlottenlund, Denmark. ICES CM 2014/ACOM:64. 147 pp.

ICES. 2016. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 183 pp.

ICES. 2019. Inter-benchmark protocol on Sole (Solea solea) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea) (IBPBrisol). ICES Scientific Reports. 1:ISS NO. 106 pp. http://doi.org/10.17895/ices.pub.4708.
Trebilcock P. and N. de Rozarieux. 2009. National Federation Fishermen's Organisation Annual Fisheries Reports. Cornish Fish Producers Organisation / Seafood Cornwall Training Ltd, March 2009.

## Table 36.1 - Sol.27.7fg - Official Nominal landings and data used by the Working Group (t)

| Year | Belgium | Denmark | France | Ireland | UK(E.\&W,NI.) | UK(Scotland) | Netherlands | TotalOfficial | Unallocated | Used by WG | TAC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 1039 * | 2 | 146 | 188 | 611 | - | 3 | 1989 | -389 | 1600 |  |
| 1987 | 701 * | - | 117 | 9 | 437 | - | - | 1264 | -42 | 1222 | 1600 |
| 1988 | 705 * | - | 110 | 72 | 317 | - | - | 1204 | -58 | 1146 | 1100 |
| 1989 | 684 * | - | 87 | 18 | 203 | - | - | 992 | 0 | 992 | 1000 |
| 1990 | 716 * | - | 130 | 40 | 353 | 0 | - | 1239 | -50 | 1189 | 1200 |
| 1991 | 982 * | - | 80 | 32 | 402 | 0 | - | 1496 | -389 | 1107 | 1200 |
| 1992 | 543 * | - | 141 | 45 | 325 | 6 | - | 1060 | -79 | 981 | 1200 |
| 1993 | 575 * | - | 108 | 51 | 285 | 11 | - | 1030 | -102 | 928 | 1100 |
| 1994 | 619 * | - | 90 | 37 | 264 | 8 | - | 1018 | -9 | 1009 | 1100 |
| 1995 | 763 * | - | 88 | 20 | 294 | - | - | 1165 | -8 | 1157 | 1100 |
| 1996 | 695 * | - | 102 | 19 | 265 | 0 | - | 1081 | -86 | 995 | 1000 |
| 1997 | 660 * | - | 99 | 28 | 251 | 0 | - | 1038 | -111 | 927 | 900 |
| 1998 | 675 * | - | 98 | 42 | 198 | - | - | 1013 | -138 | 875 | 850 |
| 1999 | 604 | - | 61 | 51 | 231 | 0 | - | 947 | 65 | 1012 | 960 |
| 2000 | 694 | - | 74 | 29 | 243 | - | - | 1040 | 51 | 1091 | 1160 |
| 2001 | 720 | - | 77 | 35 | 288 | - | - | 1120 | 48 | 1168 | 1020 |
| 2002 | 703 | - | 65 | 32 | 318 | + | - | 1118 | 227 | 1345 | 1070 |
| 2003 | 715 | - | 124 | 26 | 342 | + | - | 1207 | 185 | 1392 | 1240 |
| 2004 | 735 | - | 79 | 33 | 283 | - | - | 1130 | 119 | 1249 | 1050 |
| 2005 | 645 | - | 101 | 34 | 217 | - | - | 997 | 47 | 1044 | 1000 |
| 2006 | 576 | - | 75 | 38 | 232 | - | - | 921 | 25 | 946 | 950 |
| 2007 | 582 | - | 85 | 32 | 244 | - | - | 943 | 2 | 945 | 890 |
| 2008 | 466 | - | 68 | 28 | 218 | - | - | 780 | 20 | 800 | 964 |
| 2009 | 513 | - | 74 | 26 | 194 | - | - | 807 | -2 | 805 | 993 |
| 2010 | 620 | - | 45 | 27 | 179 | - | - | 871 | 5 | 876 | 993 |
| 2011 | 766 | - | 50 | 30 | 168 | - | - | 1013 | 16 | 1029 | 1241 |
| 2012 | 843 | - | 48 | 33 | 175 | - | - | 1099 | 5 | 1104 | 1060 |
| 2013 | 789 | - | 49 | 42 | 206 | - | - | 1086 | 6 | 1092 | 1100 |
| 2014 | 705 | - | 59 | 28 | 252 | - | - | 1044 | -2 | 1042 | 1001 |
| 2015 | 671 | - | 24 | 27 | 105 | - | - | 827 | 3 | 830 | 851 |
| 2016 | 563 | - | 72 | 21 | 175 | - | - | 831 | 0 | 831 | 779 |
| 2017 | 553 | - | 49 | 28 | 149 | - | - | 780 | -4 | 776 | 845 |
| 2018 ^ | 607 | - | 44 | 28 | 171 | - | - | 849 | 1 | 850 | 920 |

[^19]* including 7.g-k

Table 36.2 - Sol.27.7fg - Annual length distributions by fleet

| Length (cm) | UK (England \& Wales) <br> Beam trawl | $\begin{gathered} \text { Belgium } \\ \hline \text { Beam trawl } \end{gathered}$ | Ireland |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Beam trawl | Otter trawl |
| 17 |  |  |  |  |
| 18 |  |  |  |  |
| 19 |  |  |  |  |
| 20 |  |  |  | 23 |
| 21 |  | 111 | 42 | 0 |
| 22 | 48 | 954 | 70 | 349 |
| 23 | 0 | 12489 | 94 | 570 |
| 24 | 260 | 110889 | 220 | 1016 |
| 25 | 1253 | 237907 | 262 | 1141 |
| 26 | 6517 | 272917 | 828 | 1958 |
| 27 | 12725 | 277698 | 1750 | 2501 |
| 28 | 17528 | 256708 | 1816 | 5044 |
| 29 | 27843 | 195191 | 2181 | 5417 |
| 30 | 34922 | 193999 | 2827 | 4772 |
| 31 | 44484 | 137659 | 2378 | 6029 |
| 32 | 44758 | 114111 | 2804 | 4836 |
| 33 | 41430 | 88722 | 2509 | 5659 |
| 34 | 31001 | 72160 | 1685 | 3983 |
| 35 | 27989 | 61220 | 1306 | 2751 |
| 36 | 23573 | 40042 | 959 | 2419 |
| 37 | 19241 | 31740 | 959 | 2035 |
| 38 | 14965 | 26441 | 903 | 1072 |
| 39 | 14584 | 20839 | 403 | 581 |
| 40 | 10771 | 11603 | 351 | 549 |
| 41 | 6636 | 9195 | 248 | 723 |
| 42 | 4480 | 10049 | 183 | 341 |
| 43 | 3278 | 4721 | 126 | 210 |
| 44 | 1652 | 3943 | 37 | 0 |
| 45 | 1137 | 1908 | 56 | 12 |
| 46 | 674 | 1268 | 37 | 141 |
| 47 | 292 | 1303 | 14 | 0 |
| 48 | 287 | 544 | 13 | 12 |
| 49 | 75 | 131 |  | 12 |
| 50 | 52 | 571 |  | 0 |
| 51 | 52 | 108 |  | 12 |
| 52 |  |  |  |  |
| 53 |  |  |  |  |
| 54 |  |  |  |  |
| 55 |  |  |  |  |
| 56 |  |  |  |  |
| 57 |  |  |  |  |
| 58 |  |  |  |  |
| 59 |  |  |  |  |
| 60 |  |  |  |  |
| Total | 392506 | 2197139 | 25063 | 54166 |


| Table 36.3-Sol.27.7fg-Catch numbers at age (in thousands) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age/Year | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 386 | 541 | 364 | 155 | 119 | 312 | 314 | 317 | 328 | 657 | 602 |
| 3 | 270 | 903 | 1883 | 438 | 287 | 833 | 438 | 739 | 561 | 971 | 675 |
| 4 | 1341 | 314 | 748 | 863 | 336 | 559 | 349 | 338 | 748 | 875 | 792 |
| 5 | 625 | 671 | 305 | 411 | 638 | 610 | 271 | 154 | 208 | 584 | 399 |
| 6 | 433 | 329 | 352 | 209 | 304 | 558 | 244 | 159 | 154 | 180 | 377 |
| 7 | 537 | 213 | 119 | 239 | 110 | 261 | 404 | 99 | 197 | 62 | 150 |
| 8 | 763 | 232 | 110 | 97 | 102 | 131 | 120 | 198 | 124 | 96 | 120 |
| 9 | 376 | 314 | 116 | 109 | 67 | 197 | 28 | 71 | 153 | 100 | 94 |
| +gp | 1220 | 731 | 644 | 541 | 372 | 462 | 365 | 174 | 169 | 352 | 380 |
| Age/Year | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 342 | 647 | 671 | 196 | 494 | 318 | 526 | 479 | 277 | 1458 | 433 |
| 3 | 830 | 1078 | 845 | 1475 | 1296 | 958 | 464 | 1163 | 993 | 690 | 1699 |
| 4 | 309 | 729 | 605 | 767 | 1173 | 798 | 878 | 601 | 1175 | 658 | 644 |
| 5 | 467 | 284 | 541 | 566 | 526 | 578 | 441 | 621 | 399 | 496 | 409 |
| 6 | 280 | 349 | 184 | 296 | 358 | 273 | 387 | 237 | 452 | 151 | 253 |
| 7 | 207 | 225 | 277 | 100 | 193 | 205 | 127 | 188 | 138 | 156 | 61 |
| 8 | 92 | 192 | 106 | 140 | 87 | 100 | 78 | 82 | 115 | 55 | 59 |
| 9 | 111 | 52 | 47 | 73 | 103 | 61 | 67 | 24 | 50 | 46 | 28 |
| +gp | 326 | 320 | 274 | 240 | 328 | 179 | 268 | 102 | 129 | 162 | 89 |
| Age/Year | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 354 | 295 | 129 | 177 | 245 | 197 | 608 | 1721 | 701 | 29 | 132 |
| 3 | 862 | 790 | 1154 | 1036 | 890 | 931 | 1719 | 1480 | 1909 | 1465 | 776 |
| 4 | 1103 | 739 | 1096 | 905 | 599 | 724 | 834 | 683 | 856 | 2202 | 1262 |
| 5 | 332 | 864 | 419 | 424 | 400 | 297 | 282 | 241 | 434 | 660 | 2070 |
| 6 | 186 | 283 | 482 | 229 | 252 | 171 | 143 | 60 | 241 | 249 | 448 |
| 7 | 161 | 149 | 133 | 192 | 127 | 108 | 80 | 56 | 65 | 95 | 248 |
| 8 | 63 | 65 | 112 | 57 | 126 | 51 | 31 | 43 | 39 | 54 | 89 |
| 9 | 83 | 42 | 65 | 43 | 45 | 52 | 23 | 19 | 26 | 36 | 29 |
| +gp | 99 | 146 | 109 | 106 | 106 | 87 | 44 | 51 | 81 | 51 | 84 |
| Age/Year | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 476 | 290 | 684 | 335 | 214 | 607 | 281 | 124 | 160 | 436 | 115 |
| 3 | 1927 | 917 | 1329 | 865 | 452 | 464 | 1316 | 1013 | 233 | 1065 | 629 |
| 4 | 886 | 897 | 714 | 743 | 559 | 426 | 744 | 1443 | 1029 | 343 | 743 |
| 5 | 889 | 508 | 576 | 474 | 565 | 346 | 347 | 398 | 1308 | 837 | 217 |
| 6 | 807 | 426 | 163 | 325 | 277 | 292 | 258 | 273 | 364 | 693 | 430 |
| 7 | 128 | 373 | 148 | 157 | 198 | 173 | 164 | 194 | 207 | 227 | 421 |
| 8 | 67 | 51 | 178 | 145 | 76 | 103 | 118 | 133 | 136 | 80 | 138 |
| 9 | 38 | 44 | 44 | 184 | 109 | 44 | 66 | 66 | 91 | 66 | 84 |
| +gp | 55 | 45 | 51 | 70 | 172 | 193 | 118 | 199 | 246 | 166 | 218 |
| Age/Year | 2015 | 2016 | 2017 | 2018 |  |  |  |  |  |  |  |
| 1 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 2 | 85 | 514 | 228 | 336 |  |  |  |  |  |  |  |
| 3 | 806 | 428 | 1067 | 735 |  |  |  |  |  |  |  |
| 4 | 863 | 607 | 478 | 980 |  |  |  |  |  |  |  |
| 5 | 382 | 663 | 325 | 272 |  |  |  |  |  |  |  |
| 6 | 140 | 245 | 326 | 283 |  |  |  |  |  |  |  |
| 7 | 217 | 86 | 130 | 173 |  |  |  |  |  |  |  |
| 8 | 117 | 143 | 43 | 88 |  |  |  |  |  |  |  |
| 9 | 82 | 97 | 68 | 36 |  |  |  |  |  |  |  |
| + gp | 132 | 93 | 111 | 134 |  |  |  |  |  |  |  |


| Table 36.4 - Sol.27.7fg - Catch weights at age (kg) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age/Year | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 |
| 1 | 0.039 | 0.106 | 0.081 | 0.063 | 0.046 | 0.114 | 0.098 | 0.068 | 0.023 | 0.048 |
| 2 | 0.106 | 0.147 | 0.143 | 0.137 | 0.132 | 0.167 | 0.169 | 0.154 | 0.132 | 0.144 |
| 3 | 0.167 | 0.186 | 0.202 | 0.205 | 0.212 | 0.218 | 0.235 | 0.234 | 0.232 | 0.234 |
| 4 | 0.222 | 0.226 | 0.258 | 0.270 | 0.286 | 0.268 | 0.297 | 0.309 | 0.321 | 0.316 |
| 5 | 0.272 | 0.264 | 0.311 | 0.329 | 0.355 | 0.316 | 0.355 | 0.378 | 0.401 | 0.392 |
| 6 | 0.315 | 0.302 | 0.361 | 0.385 | 0.417 | 0.363 | 0.409 | 0.441 | 0.471 | 0.461 |
| 7 | 0.352 | 0.340 | 0.408 | 0.436 | 0.473 | 0.409 | 0.460 | 0.499 | 0.531 | 0.523 |
| 8 | 0.383 | 0.376 | 0.452 | 0.483 | 0.523 | 0.453 | 0.506 | 0.551 | 0.581 | 0.579 |
| 9 | 0.408 | 0.413 | 0.493 | 0.525 | 0.567 | 0.496 | 0.548 | 0.598 | 0.622 | 0.627 |
| +gp | 0.440 | 0.538 | 0.602 | 0.624 | 0.672 | 0.665 | 0.668 | 0.720 | 0.664 | 0.720 |
| Age/Year | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| 1 | 0.078 | 0.061 | 0.085 | 0.019 | 0.089 | 0.046 | 0.048 | 0.074 | 0.013 | 0.049 |
| 2 | 0.154 | 0.156 | 0.173 | 0.131 | 0.170 | 0.144 | 0.146 | 0.157 | 0.109 | 0.134 |
| 3 | 0.225 | 0.243 | 0.255 | 0.235 | 0.246 | 0.236 | 0.236 | 0.235 | 0.198 | 0.214 |
| 4 | 0.292 | 0.324 | 0.330 | 0.330 | 0.317 | 0.321 | 0.320 | 0.309 | 0.280 | 0.291 |
| 5 | 0.355 | 0.397 | 0.398 | 0.416 | 0.383 | 0.400 | 0.396 | 0.378 | 0.355 | 0.363 |
| 6 | 0.414 | 0.462 | 0.459 | 0.494 | 0.444 | 0.471 | 0.466 | 0.442 | 0.424 | 0.430 |
| 7 | 0.469 | 0.521 | 0.514 | 0.562 | 0.500 | 0.536 | 0.528 | 0.502 | 0.487 | 0.494 |
| 8 | 0.519 | 0.572 | 0.561 | 0.622 | 0.552 | 0.594 | 0.584 | 0.557 | 0.543 | 0.553 |
| 9 | 0.565 | 0.617 | 0.602 | 0.673 | 0.598 | 0.645 | 0.632 | 0.608 | 0.592 | 0.609 |
| +gp | 0.665 | 0.704 | 0.679 | 0.772 | 0.703 | 0.748 | 0.740 | 0.738 | 0.691 | 0.747 |
| Age/Year | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 1 | 0.054 | 0.073 | 0.057 | 0.081 | 0.068 | 0.027 | 0.074 | 0.079 | 0.015 | 0.078 |
| 2 | 0.150 | 0.147 | 0.134 | 0.151 | 0.147 | 0.124 | 0.156 | 0.163 | 0.122 | 0.166 |
| 3 | 0.239 | 0.216 | 0.207 | 0.216 | 0.220 | 0.214 | 0.234 | 0.244 | 0.222 | 0.248 |
| 4 | 0.320 | 0.281 | 0.275 | 0.276 | 0.288 | 0.296 | 0.307 | 0.320 | 0.315 | 0.322 |
| 5 | 0.393 | 0.342 | 0.338 | 0.331 | 0.351 | 0.372 | 0.376 | 0.393 | 0.400 | 0.390 |
| 6 | 0.459 | 0.398 | 0.396 | 0.380 | 0.409 | 0.439 | 0.440 | 0.462 | 0.478 | 0.451 |
| 7 | 0.516 | 0.451 | 0.450 | 0.425 | 0.462 | 0.500 | 0.500 | 0.528 | 0.549 | 0.506 |
| 8 | 0.566 | 0.499 | 0.500 | 0.465 | 0.510 | 0.552 | 0.555 | 0.589 | 0.613 | 0.553 |
| 9 | 0.608 | 0.543 | 0.545 | 0.500 | 0.553 | 0.598 | 0.605 | 0.647 | 0.670 | 0.594 |
| +gp | 0.674 | 0.640 | 0.645 | 0.563 | 0.643 | 0.677 | 0.707 | 0.781 | 0.765 | 0.665 |
| Age/Year | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 1 | 0.066 | 0.054 | 0.123 | 0.066 | 0.068 | 0.085 | 0.075 | 0.128 | 0.128 | 0.127 |
| 2 | 0.148 | 0.130 | 0.171 | 0.130 | 0.145 | 0.139 | 0.139 | 0.164 | 0.179 | 0.160 |
| 3 | 0.225 | 0.202 | 0.218 | 0.194 | 0.219 | 0.192 | 0.200 | 0.198 | 0.221 | 0.186 |
| 4 | 0.296 | 0.271 | 0.266 | 0.256 | 0.288 | 0.245 | 0.258 | 0.258 | 0.252 | 0.230 |
| 5 | 0.363 | 0.336 | 0.313 | 0.317 | 0.354 | 0.297 | 0.313 | 0.309 | 0.320 | 0.310 |
| 6 | 0.425 | 0.399 | 0.361 | 0.377 | 0.415 | 0.349 | 0.365 | 0.305 | 0.394 | 0.346 |
| 7 | 0.482 | 0.457 | 0.408 | 0.435 | 0.473 | 0.400 | 0.414 | 0.412 | 0.417 | 0.404 |
| 8 | 0.533 | 0.513 | 0.454 | 0.493 | 0.528 | 0.451 | 0.460 | 0.521 | 0.463 | 0.404 |
| 9 | 0.579 | 0.564 | 0.501 | 0.549 | 0.578 | 0.501 | 0.503 | 0.532 | 0.481 | 0.530 |
| +gp | 0.677 | 0.704 | 0.639 | 0.721 | 0.690 | 0.618 | 0.609 | 0.536 | 0.622 | 0.591 |
| Age/Year | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |  |  |
| 1 | 0.140 | 0.110 | 0.125 | 0.073 | 0.134 | 0.130 | 0.11 | 0.124 |  |  |
| 2 | 0.162 | 0.162 | 0.179 | 0.170 | 0.163 | 0.187 | 0.181 | 0.162 |  |  |
| 3 | 0.184 | 0.213 | 0.205 | 0.208 | 0.200 | 0.211 | 0.216 | 0.208 |  |  |
| 4 | 0.223 | 0.247 | 0.253 | 0.273 | 0.254 | 0.262 | 0.263 | 0.258 |  |  |
| 5 | 0.272 | 0.279 | 0.285 | 0.366 | 0.319 | 0.293 | 0.323 | 0.303 |  |  |
| 6 | 0.354 | 0.324 | 0.334 | 0.393 | 0.352 | 0.353 | 0.353 | 0.347 |  |  |
| 7 | 0.420 | 0.341 | 0.350 | 0.425 | 0.443 | 0.462 | 0.394 | 0.398 |  |  |
| 8 | 0.447 | 0.377 | 0.475 | 0.484 | 0.516 | 0.434 | 0.504 | 0.485 |  |  |
| 9 | 0.475 | 0.409 | 0.412 | 0.530 | 0.436 | 0.476 | 0.468 | 0.483 |  |  |
| +gp | 0.622 | 0.538 | 0.576 | 0.685 | 0.549 | 0.604 | 0.484 | 0.568 |  |  |


| Table 36.5-Sol.27.7fg - Stock weights at age (kg) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age/Year | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 |
| 1 | 0.090 | 0.090 | 0.090 | 0.090 | 0.090 | 0.090 | 0.090 | 0.090 | 0.090 | 0.090 | 0.090 | 0.090 | 0.090 |
| 2 | 0.076 | 0.113 | 0.113 | 0.113 | 0.113 | 0.113 | 0.145 | 0.113 | 0.113 | 0.113 | 0.113 | 0.113 | 0.113 |
| 3 | 0.136 | 0.157 | 0.142 | 0.159 | 0.141 | 0.160 | 0.174 | 0.167 | 0.163 | 0.157 | 0.159 | 0.164 | 0.175 |
| 4 | 0.190 | 0.222 | 0.203 | 0.221 | 0.215 | 0.210 | 0.236 | 0.257 | 0.255 | 0.238 | 0.232 | 0.255 | 0.262 |
| 5 | 0.239 | 0.298 | 0.263 | 0.305 | 0.295 | 0.269 | 0.366 | 0.360 | 0.392 | 0.354 | 0.306 | 0.356 | 0.370 |
| 6 | 0.406 | 0.351 | 0.334 | 0.450 | 0.353 | 0.354 | 0.392 | 0.413 | 0.437 | 0.394 | 0.385 | 0.487 | 0.488 |
| 7 | 0.472 | 0.352 | 0.322 | 0.448 | 0.593 | 0.432 | 0.454 | 0.521 | 0.485 | 0.622 | 0.462 | 0.543 | 0.633 |
| 8 | 0.389 | 0.593 | 0.400 | 0.464 | 0.423 | 0.462 | 0.505 | 0.508 | 0.595 | 0.556 | 0.551 | 0.610 | 0.606 |
| 9 | 0.346 | 0.417 | 0.539 | 0.624 | 0.465 | 0.425 | 0.907 | 0.560 | 0.657 | 0.704 | 0.737 | 0.766 | 0.464 |
| +gp | 0.625 | 0.588 | 0.613 | 0.654 | 0.666 | 0.620 | 0.725 | 0.723 | 0.692 | 0.737 | 0.674 | 0.799 | 0.827 |
| Age/Year | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 |
| 2 | 0.118 | 0.113 | 0.113 | 0.113 | 0.113 | 0.113 | 0.113 | 0.113 | 0.113 | 0.148 | 0.113 | 0.113 | 0.104 |
| 3 | 0.173 | 0.175 | 0.18 | 0.153 | 0.158 | 0.152 | 0.164 | 0.179 | 0.184 | 0.196 | 0.135 | 0.143 | 0.186 |
| 4 | 0.274 | 0.268 | 0.273 | 0.242 | 0.233 | 0.227 | 0.247 | 0.23 | 0.265 | 0.267 | 0.227 | 0.233 | 0.284 |
| 5 | 0.429 | 0.472 | 0.398 | 0.361 | 0.363 | 0.308 | 0.369 | 0.356 | 0.388 | 0.392 | 0.329 | 0.335 | 0.387 |
| 6 | 0.517 | 0.433 | 0.462 | 0.473 | 0.466 | 0.465 | 0.476 | 0.536 | 0.498 | 0.47 | 0.43 | 0.441 | 0.486 |
| 7 | 0.641 | 0.462 | 0.546 | 0.468 | 0.687 | 0.546 | 0.523 | 0.376 | 0.751 | 0.492 | 0.521 | 0.54 | 0.573 |
| 8 | 0.613 | 0.48 | 0.636 | 0.587 | 0.687 | 0.526 | 0.753 | 0.859 | 0.754 | 0.576 | 0.599 | 0.629 | 0.647 |
| 9 | 0.836 | 0.944 | 0.89 | 0.82 | 0.676 | 0.542 | 0.847 | 0.735 | 0.475 | 0.636 | 0.661 | 0.705 | 0.708 |
| +gp | 0.876 | 0.844 | 0.861 | 0.833 | 0.799 | 0.745 | 0.917 | 0.779 | 0.869 | 0.717 | 0.775 | 0.866 | 0.822 |
| Age/Year | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 1 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.108 | 0.115 |
| 2 | 0.113 | 0.113 | 0.11 | 0.062 | 0.113 | 0.113 | 0.158 | 0.116 | 0.149 | 0.143 | 0.117 | 0.141 | 0.151 |
| 3 | 0.178 | 0.195 | 0.204 | 0.169 | 0.187 | 0.189 | 0.205 | 0.176 | 0.213 | 0.188 | 0.177 | 0.176 | 0.19 |
| 4 | 0.276 | 0.282 | 0.317 | 0.306 | 0.312 | 0.289 | 0.258 | 0.248 | 0.275 | 0.235 | 0.236 | 0.232 | 0.223 |
| 5 | 0.386 | 0.371 | 0.433 | 0.434 | 0.434 | 0.403 | 0.317 | 0.329 | 0.337 | 0.284 | 0.294 | 0.274 | 0.287 |
| 6 | 0.495 | 0.454 | 0.541 | 0.534 | 0.538 | 0.512 | 0.381 | 0.415 | 0.399 | 0.334 | 0.35 | 0.261 | 0.349 |
| 7 | 0.598 | 0.529 | 0.635 | 0.603 | 0.619 | 0.609 | 0.449 | 0.502 | 0.459 | 0.386 | 0.406 | 0.389 | 0.357 |
| 8 | 0.689 | 0.593 | 0.712 | 0.648 | 0.68 | 0.691 | 0.521 | 0.587 | 0.52 | 0.441 | 0.46 | 0.542 | 0.437 |
| 9 | 0.766 | 0.644 | 0.772 | 0.677 | 0.725 | 0.757 | 0.594 | 0.667 | 0.579 | 0.496 | 0.513 | 0.526 | 0.501 |
| +gp | 0.926 | 0.750 | 0.878 | 0.709 | 0.790 | 0.881 | 0.850 | 0.886 | 0.782 | 0.709 | 0.687 | 0.564 | 0.620 |
| Age/Year | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |  |  |  |  |
| 1 | 0.112 | 0.13 | 0.086 | 0.107 | 0.049 | 0.113 | 0.11 | 0.091 | 0.115 |  |  |  |  |
| 2 | 0.143 | 0.143 | 0.151 | 0.14 | 0.146 | 0.109 | 0.158 | 0.153 | 0.134 |  |  |  |  |
| 3 | 0.183 | 0.172 | 0.186 | 0.182 | 0.193 | 0.184 | 0.186 | 0.201 | 0.194 |  |  |  |  |
| 4 | 0.226 | 0.204 | 0.213 | 0.232 | 0.237 | 0.23 | 0.229 | 0.236 | 0.236 |  |  |  |  |
| 5 | 0.28 | 0.25 | 0.249 | 0.265 | 0.304 | 0.295 | 0.273 | 0.291 | 0.282 |  |  |  |  |
| 6 | 0.333 | 0.331 | 0.297 | 0.305 | 0.335 | 0.359 | 0.336 | 0.322 | 0.335 |  |  |  |  |
| 7 | 0.399 | 0.381 | 0.347 | 0.337 | 0.377 | 0.417 | 0.403 | 0.373 | 0.375 |  |  |  |  |
| 8 | 0.41 | 0.425 | 0.398 | 0.403 | 0.412 | 0.468 | 0.439 | 0.483 | 0.437 |  |  |  |  |
| 9 | 0.495 | 0.438 | 0.428 | 0.394 | 0.502 | 0.459 | 0.496 | 0.451 | 0.493 |  |  |  |  |
| + gp | 0.615 | 0.638 | 0.583 | 0.554 | 0.640 | 0.598 | 0.580 | 0.574 | 0.596 |  |  |  |  |

Table 36.6-Sol.27.7fg - Discard rates



Table 36.8 - Sol. 27.7 fg - Indices of effort.

| Year | England \& Wales |  |  |  | Belgium |  | Ireland |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Otter trawl ${ }^{1}$ | Otter traw ${ }^{6}$ | Beam trawl ${ }^{1}$ | Beam traw ${ }^{6}$ | Beam trawl ${ }^{3}$ | Beam traw ${ }^{5}$ | Otter traw ${ }^{4}$ | Scottish seine ${ }^{5}$ | Beam traw $1^{5}$ |
| 1971 | - | - | - | - | 11.06 | - | - | - | - |
| 1972 | 45.72 | - | - | - | 8.44 | - | - | - | - |
| 1973 | 45.28 | - | - | - | 17.39 | - | - | - | - |
| 1974 | 38.94 | - | - | - | 18.83 | - | - | - | - |
| 1975 | 33.53 | - | - | - | 16.38 | - | - | - | - |
| 1976 | 25.61 | - | - | - | 28.07 | - | - | - | - |
| 1977 | 27.16 | - | - | - | 24.11 | - | - | - | - |
| 1978 | 27.08 | - | 2.50 | - | 18.09 | - | - | - | - |
| 1979 | 23.84 | - | 1.96 | - | 18.90 | - | - | - | - |
| 1980 | 26.43 | - | 4.31 | - | 29.02 | - | - | - | - |
| 1981 | 24.10 | - | 6.24 | - | 35.39 | - | - | - | - |
| 1982 | 19.20 | - | 9.95 | - | 28.77 | - | - | - | - |
| 1983 | 17.61 | 620 | 12.35 | 195 | 34.95 | - | - | - | - |
| 1984 | 23.16 | 1723 | 13.55 | 901 | 33.48 | - | - | - | - |
| 1985 | 25.24 | 1493 | 18.70 | 1101 | 40.49 | - | - | - | - |
| 1986 | 21.18 | 1125 | 20.72 | 973 | 52.46 | - | - | - | - |
| 1987 | 24.43 | 1211 | 38.76 | 1681 | 37.26 | - | - | - | - |
| 1988 | 20.09 | 838 | 25.62 | 1102 | 42.92 | - | - | - | - |
| 1989 | 17.61 | 966 | 20.26 | 861 | 53.58 | - | - | - | - |
| 1990 | 22.56 | 1229 | 30.77 | 1256 | 40.27 | - | - | - | - |
| 1991 | 18.57 | 1066 | 40.81 | 1667 | 18.05 | - | - | - | . |
| 1992 | 16.00 | 898 | 35.78 | 1420 | 25.47 | - | - | - | - |
| 1993 | 13.79 | 836 | 39.64 | 1669 | 31.27 | - | - | - | - |
| 1994 | 9.48 | 623 | 37.03 | 2219 | 38.35 | - | - | - | - |
| 1995 | 8.46 | 580 | 37.59 | 2303 | 47.81 | - | 63.33 | 6.43 | 20.69 |
| 1996 | 8.67 | 593 | 39.78 | 2391 | 47.63 | 53.27 | 59.97 | 9.73 | 26.70 |
| 1997 | 8.14 | 577 | 43.00 | 2661 | 51.98 | 57.36 | 65.00 | 16.07 | 28.16 |
| 1998 | 7.13 | 517 | 47.84 | 2846 | 52.11 | 57.79 | 72.25 | 14.88 | 35.33 |
| 1999 | 5.69 | 395 | 50.87 | 3058 | 55.03 | 55.11 | 51.48 | 8.01 | 41.04 |
| 2000 | 4.05 | 284 | 51.19 | 3133 | 56.05 | 51.34 | 60.56 | 9.86 | 36.91 |
| 2001 | 4.42 | 309 | 49.32 | 3172 | 52.06 | 54.90 | 69.37 | 16.33 | 39.50 |
| 2002 | 6.10 | 416 | 37.53 | 2652 | 43.24 | 49.60 | 77.20 | 20.88 | 31.49 |
| 2003 | 9.94 | 696 | 40.71 | 2669 | 42.81 | 62.73 | 86.78 | 20.07 | 49.39 |
| 2004 | 9.42 | 641 | 32.37 | 2503 | - | 78.73 | 97.12 | 18.42 | 57.77 |
| 2005 | 12.09 | 876 | 27.73 | 1968 | - | 64.50 | 124.67 | 14.64 | 51.67 |
| 2006 | 12.97 | 924 | 18.57 | 1330 | - | 49.61 | 118.04 | 14.78 | 63.21 |
| 2007 | 10.66 | 798 | 15.37 | 1407 | - | 45.91 | 135.36 | 15.81 | 56.59 |
| 2008 | 10.13 | 711 | 13.83 | 1202 | - | 28.72 | 125.41 | 11.65 | 38.66 |
| 2009 | 8.97 | 656 | 12.31 | 1105 | - | 30.65 | 137.11 | 8.18 | 39.11 |
| 2010 | 7.67 | 565 | 14.44 | 1162 | - | 32.46 | 140.79 | 9.68 | 40.97 |
| 2011 | 7.44 | 525 | 13.79 | 868 | - | 38.77 | 120.33 | 11.05 | 36.07 |
| 2012 | 7.79 | 543 | 12.77 | 1408 | - | 46.25 | 127.68 | 14.21 | 40.49 |
| 2013 | 4.27 | 280 | 0.78 | 1611 | - | 45.23 | 118.20 | 13.15 | 38.74 |
| 2014 | - | 156 | - | 959 | - | 31.30 | 127.34 | 12.46 | 37.88 |
| 2015 | - | 79 | - | 726 | - | 31.79 | 132.69 | 9.29 | 37.79 |
| 2016 | - | 0 | - | 915 | - | 32.34 | 148.17 | 10.44 | 39.67 |
| 2017 | - | 93 | - | 986 | - | 33.35 | 136.05 | 9.76 | 35.24 |
| 2018* | - | 127 | - | 1071 | - | 31.48 | 108.22 | 9.69 | 37.44 |
| ${ }^{1}$ Division Vlif only - Fishing hours (x10^3) corrected for fishing power |  |  |  |  |  |  |  |  |  |
| ${ }^{2}$ Days at sea Vlifg |  |  |  |  |  |  |  |  |  |
| ${ }^{3}$ Fis hing hours ( $\mathrm{x} 10^{\wedge} 3$ ) corrected for fishing power using P $=0.000204 \mathrm{BHP}$ ^1.23 |  |  |  |  |  |  |  |  |  |
| ${ }^{4}$ Division Vlig only - Fishing hours ( $\times 10^{\wedge} 3$ ) |  |  |  |  |  |  |  |  |  |
| ${ }^{5}$ Fishing hours ( $\times 10^{\wedge} 3$ ) |  |  |  |  |  |  |  |  |  |
| ${ }^{6}$ Division Vlif only - Days fished corrected for fishing power * provisional |  |  |  |  |  |  |  |  |  |


| Table 36.9-Sol.27.7fg - LPUE |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | UK | England \& Wales |  |  |  |  |  | Belgium |  | Ireland |  |  |
|  | BT Surve ${ }^{4}$ | Otter traw ${ }^{1}$ | Otter trawl ${ }^{6}$ | Otter trawi ${ }^{1}$ | Otter traw ${ }^{6}$ | Beam trawi ${ }^{1}$ | Beam trawl ${ }^{6}$ | Beam traw ${ }^{2}$ | Beam traw ${ }^{5}$ | Otter traw ${ }^{5}$ | Scottish sein ${ }^{5}$ | Beam trawl ${ }^{5}$ |
|  | Division Vllfg | Division Vllf | Division VIIf | Division $\mathrm{VIIg}^{3}$ | Division $\mathrm{VlIg}^{3}$ | Division VIIf | Division VIIf | Division Vllfg | Division VIIfg | Division VIIIg | Division VIlg | Division Vllg |
| 1971 | - | - | - | - | - | - | - | 47.92 | - | - | - | - |
| 1972 | - | 2.42 | - | 2.11 | - | - | - | 37.06 | - | - | - | - |
| 1973 | - | 2.45 | - | 0.98 | - | - | - | 39.47 | - | - | - | - |
| 1974 | - | 2.10 | - | 1.83 | - | - | - | 37.81 | - | - | - | - |
| 1975 | - | 1.82 | - | 1.79 | - | - | - | 31.41 | - | - | - | - |
| 1976 | - | 2.02 | - | 1.30 | - | - | - | 30.50 | - | - | - | - |
| 1977 | - | 1.84 | - | 1.21 | - | - | - | 27.90 | - | - | - | - |
| 1978 | - | 1.82 | - | 1.17 | - | 13.99 | - | 23.35 | - | - | - | - |
| 1979 | - | 1.80 | - | 1.15 | - | 14.83 | - | 33.19 | - | - | - | - |
| 1980 | - | 1.86 | - | 1.55 | - | 18.99 | - | 29.73 | - | - | - | - |
| 1981 | - | 1.45 | - | 0.60 | - | 13.58 | - | 24.03 | - | - | - | - |
| 1982 | - | 1.73 | - | 0.56 | - | 11.79 | - | 25.93 | - | - | - | - |
| 1983 | - | 2.22 | 30.54 | 1.14 | 35.75 | 13.50 | 201.80 | 22.18 | - | - | - | - |
| 1984 | - | 1.53 | 19.53 | 1.70 | 28.04 | 13.59 | 204.65 | 20.78 | - | - | - | - |
| 1985 | - | 1.55 | 26.58 | 1.55 | 37.31 | 12.52 | 240.45 | 17.94 | - | - | - | - |
| 1986 | - | 1.38 | 25.55 | 0.99 | 21.27 | 10.94 | 247.74 | 17.83 | - | - | - | - |
| 1987 | - | 0.94 | 19.85 | 1.15 | 36.02 | 7.31 | 179.34 | 17.32 | - | - | - | - |
| 1988 | 79.52 | 0.62 | 11.13 | 0.27 | 8.88 | 4.39 | 110.35 | 15.29 | - | - | - | - |
| 1989 | 150.02 | 0.99 | 17.36 | 0.87 | 18.75 | 5.38 | 130.42 | 11.33 | - | - | - | - |
| 1990 | 93.61 | 0.76 | 13.41 | 0.67 | 18.08 | 5.98 | 148.47 | 15.64 | - | - | - | - |
| 1991 | 122.06 | 0.69 | 12.26 | 0.85 | 16.20 | 4.80 | 119.52 | 24.24 | - | - | - | - |
| 1992 | 121.41 | 1.00 | 17.90 | 1.25 | 20.99 | 4.14 | 105.84 | 18.57 | - | - | - | - |
| 1993 | 76.37 | 0.55 | 8.85 | 0.25 | 4.27 | 4.80 | 118.08 | 15.21 | - | - | - | - |
| 1994 | 109.74 | 0.90 | 13.00 | 0.27 | 3.50 | 4.26 | 70.00 | 13.94 | - | - | - | - |
| 1995 | 69.91 | 0.96 | 13.76 | 0.87 | 12.75 | 4.52 | 73.20 | 13.62 | - | 0.40 | 0.62 | 0.81 |
| 1996 | 71.71 | 0.66 | 9.69 | 0.52 | 6.95 | 3.94 | 65.05 | 11.27 | 11.45 | 0.73 | 0.05 | 0.88 |
| 1997 | 81.67 | 0.86 | 12.55 | 0.52 | 6.42 | 3.28 | 53.81 | 9.96 | 9.68 | 0.42 | 0.23 | 1.16 |
| 1998 | 137.11 | 0.60 | 8.24 | 0.40 | 4.85 | 2.67 | 44.86 | 10.12 | 9.64 | 0.48 | 0.11 | 1.11 |
| 1999 | 168.46 | 0.91 | 13.25 | 0.74 | 8.18 | 3.21 | 52.36 | 11.26 | 12.14 | 0.17 | 0.09 | 0.50 |
| 2000 | 228.46 | 0.49 | 7.01 | 1.85 | 23.26 | 3.36 | 53.85 | 11.90 | 13.77 | 0.19 | 0.05 | 0.26 |
| 2001 | 158.08 | 1.14 | 17.1 | 2.13 | 27.5 | 4.02 | 62.39 | 13.25 | 13.60 | 0.31 | 0.55 | 0.18 |
| 2002 | 121.89 | 0.78 | 11.61 | 3.60 | 47.01 | 5.64 | 79.47 | 18.71 | 17.80 | 0.43 | 0.29 | 0.14 |
| 2003 | 123.91 | 0.57 | 8.03 | 0.00 | 0.00 | 5.23 | 80.85 | 19.48 | 11.40 | 0.12 | 0.03 | 0.19 |
| 2004 | 152.03 | 0.60 | 8.84 | 0.19 | 2.70 | 5.75 | 76.09 | - | 9.17 | 0.19 | 0.02 | 0.20 |
| 2005 | 76.28 | 0.76 | 10.67 | 0.26 | 3.07 | 4.94 | 70.02 | - | 9.78 | 0.14 | 0.00 | 0.29 |
| 2006 | 68.96 | 1.16 | 16.40 | 0.60 | 6.23 | 5.97 | 81.57 | - | 10.63 | 0.11 | 0.05 | 0.26 |
| 2007 | 80.95 | 0.78 | 10.75 | 1.00 | 15.04 | 9.87 | 92.17 | - | 11.53 | 0.13 | 0.02 | 0.20 |
| 2008 | 115.96 | 0.82 | 11.94 | 0.86 | 10.67 | 9.46 | 94.85 | - | 14.35 | 0.12 | 0.02 | 0.29 |
| 2009 | 90.64 | 0.94 | 13.13 | 0.46 | 6.88 | 6.37 | 69.37 | - | 14.01 | 0.10 | 0.00 | 0.28 |
| 2010 | 109.55 | 1.01 | 13.59 | 0.63 | 8.63 | 5.92 | 79.90 | - | 16.68 | 0.13 | 0.01 | 0.20 |
| 2011 | 99.47 | 1.47 | 20.78 | 0.31 | 4.47 | 6.72 | 109.20 | - | 17.90 | 0.19 | 0.01 | 0.20 |
| 2012 | 101.45 | 1.67 | 24.10 | 0.47 | 5.17 | 6.47 | 80.16 | - | 17.01 | 0.15 | 0.01 | 0.48 |
| 2013 | 119.38 | 1.76 | 27.81 | 0.34 | 4.62 | - | 82.82 | - | 16.54 | 0.14 | 0.01 | 0.65 |
| 2014 | 86.75 | - | 6.19 | - | 11.56 | $\bullet$ | 107.25 | - | 21.30 | 0.12 | - | 0.34 |
| 2015 | 85.45 | - | 51.13 | - | 5.62 | - | 103.07 | - | 20.14 | 0.11 | - | 0.31 |
| 2016 | 113.55 | - | 0.00 | - | 0.00 | - | 113.16 | - | 16.25 | 0.10 | 0.01 | 0.20 |
| 2017 | 111.38 | - | 31.29 | - | 18.09 | $\bullet$ | 100.03 | - | 15.72 | 0.18 | 0.05 | 0.22 |
| 2018* | 206.44 |  | 36.37 | , | 4.86 |  | 119.89 | - | 18.09 | 0.18 | - | 0.27 |
| ${ }^{1} \mathrm{Kg} / \mathrm{hr}$ corrected for GRT. <br> ${ }^{2} \mathrm{Kg} / \mathrm{hr}$ corrected for fishing power using $\mathrm{P}=0.000204 \mathrm{BHP}{ }^{\wedge} 1.23$ <br> ${ }^{3}$ Division VIIg (East). <br> ${ }^{4} \mathrm{Kg} / 100 \mathrm{~km}$ <br> ${ }^{5} \mathrm{Kg} /$ hour <br> ${ }^{6} \mathrm{Kg} /$ day <br> * provisional |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |




## Table 34.10-Sol.27.7fg -Tuning series continued

Indices in bold are used in the assessment

| IR - GFS : Irish Groundfish Survey (IBTS 4th Qtr) - 7.g Sole number at age (Interim indices for new Celtic Explorer series) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0.79 | 0.92 |  |  |  |  |  |  |  |
| 1 | 10 |  |  |  |  |  |  |  |  |  |
| 832 | 1.0 | 5.2 | 1.1 | 3.2 | 3.0 | 4.1 | 4.0 | 0.0 | 1.0 | 0.0 |
| 980 | 1.0 | 8.0 | 6.0 | 5.0 | 1.0 | 2.0 | 1.0 | 0.0 | 0.0 | 1.0 |
| 845 | 0.0 | 0.0 | 6.0 | 2.0 | 4.0 | 2.0 | 2.0 | 0.0 | 0.0 | 0.0 |
| 1046 | 0.0 | 0.0 | 4.0 | 4.0 | 6.0 | 4.0 | 1.0 | 0.0 | 0.0 | 0.0 |
| 1168 | 0.0 | 2.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 0.0 | 0.0 | 0.0 |
| 1139 | 2.0 | 9.0 | 7.0 | 3.0 | 2.0 | 0.0 | 2.0 | 0.0 | 1.0 | 0.0 |
| 1018 | 0.0 | 15.0 | 3.0 | 4.0 | 1.0 | 1.0 | 2.0 | 1.0 | 0.0 | 2.0 |
| 1381 | 0.0 | 12.0 | 24.7 | 9.1 | 8.2 | 1.0 | 3.0 | 3.9 | 0.0 | 2.1 |
| 1392 | 2.0 | 0.0 | 20.1 | 8.0 | 6.1 | 3.1 | 0.0 | 1.0 | 1.0 | 3.7 |
| 1470 | 0.0 | 7.0 | 3.0 | 3.0 | 3.0 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1439 | 0.0 | 2.0 | 10.0 | 7.5 | 1.8 | 2.0 | 3.8 | 2.0 | 1.0 | 1.0 |
| 1487 | 0.1 | 3.4 | 7.7 | 8.0 | 6.1 | 3.7 | 0.5 | 0.1 | 0.1 | 0.1 |

UK (E+W) TRAWL 107F. (Processed as unsexed - from 2001WG)
(LPUE data reprocessed in 2014. Effort changed from hours to days)


## Table 36.11. Sol.27.7fg - Diagnostics.

```
FLR XSA Diagnostics 2019-05-14 10:21:23
```

CPUE data from indices
Catch data for 48 years. 1971 to 2018. Ages 1 to 10 .

|  | fleet first age last age first | year | last | year | alpha beta |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | BE-CBT | 3 | 9 | 1971 | 1996 | 0 | 1 |
| 2 | BE-CBT3 | 2 | 9 | 2006 | 2018 | 0 | 1 |
| 3 | UK (E\&W)-CBT | 3 | 8 | 1991 | 2012 | 0 | 1 |
| 4 | UK (E\&W) - BTS-Q3 | 1 | 5 | 1988 | 2018 | 0.75 | 0.85 |

```
Time series weights :
    Tapered time weighting not applied
```

Catchability analysis :
Catchability independent of size for all ages
Catchability independent of age for ages > 7
Terminal population estimation :
Survivor estimates shrunk towards the mean F
of the final 5 years or the 5 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1.5$
Minimum standard error for population
estimates derived from each fleet $=0.3$
prior weighting not applied
Regression weights
year
$\begin{array}{crrrrrrrrr}\text { age } & 2009 & 2010 & 2011 & 2012 & 2013 & 2014 & 2015 & 2016 & 2017 \\ \text { all } & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$
Fishing mortalities
year
age $\begin{array}{lllllllllll}2009 & 2010 & 2011 & 2012 & 2013 & 2014 & 2015 & 2016 & 2017 & 2018\end{array}$
0.0000 .0000 .0000 .0000 .0000 .0000 .0000 .0000 .0000 .000
0.0740 .0490 .0740 .0400 .0830 .0280 .0330 .0660 .0390 .029
0.1520 .2020 .2260 .1730 .3570 .1480 .2470 .2060 .1710 .153
$0.2550 .3430 .317 \quad 0.335 \quad 0.368 \quad 0.4020 .2770 .2660 .3320 .210$
0.2750 .3040 .2770 .4670 .4420 .3740 .3310 .3160 .1990 .286
$\begin{array}{llllllllllllllll}0.314 & 0.302 & 0.369 & 0.390 & 0.428 & 0.380 & 0.389 & 0.326 & 0.226 & 0.239\end{array}$
$7 \quad 0.309 \quad 0.260 \quad 0.346 \quad 0.4690 .3990 .4460 .298 \quad 0.3930 .2560 .162$
$8 \quad 0.2400 .3190 .3100 .3870 .2950 .4000 .1890 .2910 .3070 .247$
$9 \quad 0.2700 .2140 .2640 .3210 .2920 .508 \quad 0.390 \quad 0.2130 .1960 .407$
$100.2700 .2140 .2640 .3210 .2920 .508 \quad 0.390 \quad 0.2130 .1960 .407$
XSA population number (Thousand)
age
$\begin{array}{lllllllllll}\text { year } & 1 & 2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10\end{array}$
$\begin{array}{lllllllllll}2009 & 6758 & 8995 & 3461 & 1986 & 1514 & 1138 & 684 & 507 & 195 & 854\end{array}$
$\begin{array}{llllllllllllllll}2010 & 2023 & 6115 & 7561 & 2690 & 1392 & 1041 & 752 & 454 & 360 & 643\end{array}$
$\begin{array}{lllllllllll}2011 & 4735 & 1830 & 5266 & 5590 & 1727 & 930 & 696 & 525 & 299 & 899\end{array}$
$\begin{array}{lllllllllllll}2012 & 6372 & 4284 & 1538 & 3802 & 3685 & 1184 & 582 & 446 & 348 & 939\end{array}$
$\begin{array}{llllllllllllllllll}2013 & 4861 & 5766 & 3724 & 1170 & 2461 & 2090 & 725 & 329 & 274 & 687\end{array}$
$20143051433848032357 \quad 73314311232440 \quad 222573$
$\begin{array}{lllllllllll}2015 & 9333 & 2760 & 3870 & 3748 & 1426 & 456 & 885 & 714 & 267 & 428\end{array}$
$\begin{array}{lllllllllll}2016 & 6969 & 8445 & 2417 & 2735 & 2570 & 927 & 280 & 595 & 535 & 512\end{array}$
20171352463067152178018981695606171402655
$\begin{array}{lllllllllllllll}2018 & 5997 & 12237 & 5489 & 5456 & 1156 & 1408 & 1223 & 424 & 114 & 422\end{array}$
Estimated population abundance at 1st Jan 2019
age

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

    2019054271075242644000786100394230068
    
## Fleet: BE-CBT

Log catchability residuals.

| year |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| age | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| 3 | -0.512 | 0.161 | 0.374 | -0.109 | -0.350 | 0.398 | 0.129 | 0.048 | 0.056 |
| 4 | 0.269 | -0.186 | 0.121 | -0.041 | -0.310 | -0.019 | 0.000 | 0.053 | 0.384 |
| 5 | 0.303 | 0.151 | 0.151 | 0.125 | 0.001 | 0.247 | -0.094 | -0.450 | 0.105 |
| 6 | 0.103 | 0.268 | -0.113 | 0.422 | 0.215 | -0.210 | 0.049 | -0.260 | 0.028 |
| 7 | 0.431 | -0.030 | -0.325 | 0.120 | 0.293 | 0.110 | 0.170 | -0.410 | 0.584 |
| 8 | 0.276 | 0.171 | -0.409 | -0.014 | -0.428 | 0.513 | -0.026 | -0.155 | 0.298 |
| 9 | 0.020 | -0.107 | -0.199 | 0.187 | -0.075 | 0.123 | -0.306 | -0.213 | 0.062 |
|  |  |  |  |  |  |  |  |  |  |
| year |  |  |  |  |  |  |  |  |  |
| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| 3 | 0.024 | 0.190 | 0.085 | -0.059 | -0.228 | -0.087 | -0.021 | -0.193 | -0.571 |
| 4 | 0.246 | -0.112 | -0.169 | -0.278 | -0.375 | -0.156 | -0.113 | -0.019 | -0.212 |
| 5 | 0.159 | -0.169 | 0.012 | -0.268 | -0.023 | 0.065 | -0.084 | -0.030 | -0.076 |
| 6 | -0.105 | 0.121 | 0.138 | -0.254 | -0.169 | -0.008 | 0.015 | 0.302 | -0.080 |
| 7 | -0.888 | 0.117 | 0.331 | 0.082 | 0.152 | -0.118 | -0.010 | 0.608 | -0.045 |
| 8 | -0.176 | -0.130 | 0.343 | 0.460 | -0.102 | 0.162 | -0.285 | -0.156 | 0.558 |
| 9 | 0.022 | 0.115 | 0.487 | -0.183 | -0.285 | -0.023 | -0.039 | 0.226 | 0.082 |
|  |  |  |  |  |  |  |  |  |  |
| year |  |  |  |  |  |  |  |  |  |
| age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |  |
| 3 | -0.511 | 0.148 | 0.387 | 0.386 | 0.250 | -0.245 | 0.050 | 0.198 |  |
| 4 | -0.168 | 0.108 | 0.059 | 0.294 | -0.068 | 0.200 | 0.367 | 0.125 |  |
| 5 | -0.136 | -0.066 | -0.019 | 0.213 | -0.215 | 0.142 | -0.006 | -0.037 |  |
| 6 | 0.047 | 0.171 | -0.396 | -0.023 | -0.395 | 0.302 | -0.115 | -0.052 |  |
| 7 | 0.166 | 0.200 | -0.466 | -0.874 | 0.228 | -0.115 | 0.090 | -0.400 |  |
| 8 | 0.150 | 0.325 | -0.300 | -0.926 | 0.472 | -0.656 | 0.021 | -0.194 |  |
| 9 | -0.215 | -0.084 | -0.140 | -0.204 | 0.442 | 0.173 | -0.010 | -0.114 |  |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean_Logq | -5.0597 | -4.8417 | -4.8479 | -4.8637 | -4.941 | -4.941 | -4.941 |
| S.E_Logq | 0.2670 | 0.2670 | 0.2670 | 0.2670 | 0.267 | 0.267 | 0.267 |

Fleet: BE-CBT3
Log catchability residuals.

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| 2 | 1.051 | 0.488 | -0.047 | 0.078 | -0.196 | 0.070 | -0.592 | 0.163 | -0.573 | -0.526 | 0.361 | -0.035 | -0.241 |
| 3 | 0.525 | 0.241 | -0.204 | -0.512 | -0.084 | -0.012 | -0.425 | 0.192 | -0.309 | 0.338 | 0.130 | 0.093 | 0.028 |
| 4 | -0.010 | -0.103 | -0.123 | -0.405 | 0.089 | -0.047 | -0.084 | -0.072 | 0.260 | 0.045 | 0.036 | 0.349 | 0.066 |
| 5 | -0.053 | -0.102 | 0.232 | -0.414 | -0.161 | -0.242 | 0.267 | 0.102 | 0.148 | 0.119 | 0.185 | -0.181 | 0.100 |
| 6 | -0.344 | -0.268 | 0.005 | -0.206 | -0.166 | 0.067 | 0.127 | 0.076 | 0.156 | 0.192 | 0.125 | 0.119 | 0.118 |
| 7 | -0.552 | 0.138 | -0.106 | -0.012 | -0.260 | 0.144 | 0.335 | 0.097 | 0.438 | 0.135 | -0.351 | 0.123 | -0.127 |
| 8 | -0.938 | -0.153 | -0.069 | -0.514 | 0.039 | -0.018 | 0.115 | -0.636 | 0.048 | -0.361 | 0.023 | -0.054 | -0.364 |
| 9 | -0.041 | -0.276 | 0.218 | -0.364 | -0.657 | -0.299 | -0.097 | -0.369 | 0.468 | 0.402 | -0.321 | -0.410 | 0.347 |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean_Logq | -9.3293 | -8.1376 | -7.811 | -7.7952 | -7.8722 | -7.9801 | -7.9801 | -7.9801 |
| S.E_Logq | 0.3000 | 0.3000 | 0.300 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |


| UK (E\&W)-CBT |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Log catchability residuals. year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | - 1997 | 71998 |  | 9920 | 00 | 2001 |
| 3 | 0.055 | 0.310 | -0.151 | -0.251 | -0.123 | 0.179 | -0.331 | $1-0.128$ | 280.2 | $29-0.1$ | 87 | -0.468 |
| 4 | 0.502 | 0.092 | -0.027 | -0.513 | -0.399 | 0.230 | 0.012 | $2-0.16$ | -0.1 | 59-0.12 | 27 | -0.689 |
| 5 | 0.529 | 0.043 | -0.092 | -0.244 | -0.266 | -0.064 | -0.045 | 50.137 | -0.0 | 84-0.2 | 46 | -0.446 |
| 6 | 0.385 | 0.156 | -0.245 | -0.380 | 0.128 | -0.063 | 0.137 | 70.007 | 0.0 | .052-0.378 | 78 | -0.317 |
| 7 | 0.364 | -0.062 | 0.086 | -0.209 | -0.149 | -0.047 | 0.028 | $8-0.322$ | $22-0.1$ | 40-0.10 | 107 | -0.355 |
| 8 | 0.522 | -0.161 | -0.306 | 0.027 | 0.490 | 0.026 | 6.248 | 80.010 | 100.0 | 870.0 | 000 | -0.129 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 |  | 2012 |
| 3 | -0.193 | -0.164 | -0.394 | -0.125 | 0.240 | 0.574 | 0.8570 | 0.351 | 0.030 | -0.093 | -0. | 215 |
| 4 | -0.105 | -0.271 | -0.527 | -0.014 | 0.255 | 0.288 | 0.7240 | 0.620 | 0.096 | 0.052 |  | 125 |
| 5 | -0.270 | 0.061 | -0.247 | -0.266 | 0.328 | 0.365 | 0.2330 | 0.571 | 0.103 | -0.236 |  | 139 |
| 6 | -0.237 | 0.044 | 0.026 | -0.346 | -0.113 | 0.537 | 0.3090 | 0.529 | 0.139 | 0.026 | -0. | . 397 |
| 7 | -0.061 | -0.119 | 0.133 | 0.187 | -0.011 | 0.419 | 0.5590 | 0.095 | 0.048 | -0.224 | -0.1 | 111 |
| 8 | 0.491 | 0.109 | 0.260 | -0.009 | 0.517 | 0.630 | 0.4690 | $0.435-0$ | -0.002 | -0.104 | 0. | . 060 |

Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

|  | 3 | 4 | 5 | 6 | 7 | 8 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean_Logq | -6.8983 | -6.2752 | -5.9454 | -5.7335 | -5.6402 | -5.6402 |
| S.E_Logq | 0.2947 | 0.2947 | 0.2947 | 0.2947 | 0.2947 | 0.2947 |
|  |  |  |  |  |  |  |

Log catchability residuals.

| year |  |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| age | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |  |
| 1 | -1.379 | -0.180 | -0.477 | -0.229 | 0.198 | -0.688 | 0.347 | -0.654 | -0.653 | 0.094 |  |
| 2 | 0.132 | 0.405 | 0.508 | 0.263 | 0.220 | 0.401 | 0.432 | 0.193 | 0.205 | -0.142 |  |
| 3 | 0.387 | 1.151 | 0.196 | 0.565 | 0.632 | 0.012 | 0.851 | 0.218 | 0.537 | -0.528 |  |
| 4 | -0.152 | 0.529 | -0.091 | 0.149 | 0.774 | -0.215 | 0.341 | -0.194 | 0.612 | 0.136 |  |
| 5 | -0.137 | 0.406 | -0.058 | 0.687 | 1.007 | -1.033 | -0.253 | 0.066 | 0.080 | 0.928 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| year |  |  |  |  |  |  |  |  |  |  |  |
| age | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |  |
| 1 | 0.530 | 0.856 | 0.469 | 0.204 | 0.302 | 0.038 | 0.567 | -0.045 | 0.033 | 0.098 |  |
| 2 | 0.344 | -0.228 | 0.647 | 0.402 | 0.018 | 0.365 | 0.359 | -0.318 | -0.173 | -0.537 |  |
| 3 | 0.245 | -0.427 | -0.620 | 0.529 | 0.453 | -0.071 | 0.261 | -0.520 | -0.415 | -0.097 |  |
| 4 | 0.122 | 0.114 | 0.212 | -0.089 | 0.540 | -0.167 | -0.292 | -0.607 | -0.454 | -0.462 |  |
| 5 | 0.630 | 0.630 | -0.163 | -0.142 | 0.243 | 0.661 | 0.363 | -0.269 | -0.618 | -0.486 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| year |  |  |  |  |  |  |  |  |  |  |  |
| age | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| 1 | 0.367 | 0.040 | -0.542 | 0.381 | 0.161 | -0.232 | 0.098 | 0.099 | 0.128 | 0.066 | 0.000 |
| 2 | -0.816 | -0.489 | 0.269 | -0.702 | -0.094 | 0.450 | -0.022 | -1.230 | -0.310 | -0.766 | 0.214 |
| 3 | -1.169 | -0.804 | 0.057 | 0.568 | -1.146 | -0.148 | -0.002 | -0.256 | -1.130 | -0.500 | 1.170 |
| 4 | -0.238 | -0.955 | -0.272 | 0.386 | 0.523 | 0.357 | -0.316 | -0.041 | 0.119 | -0.156 | -0.111 |
| 5 | -0.708 | -0.240 | -1.742 | -0.593 | 0.054 | 0.473 | 0.054 | -0.359 | -0.330 | 0.115 | 0.732 |

Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

|  | 1 | 2 | 3 | 4 | 5 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Mean_Logq | -7.1506 | -7.3196 | -8.5522 | -9.0505 | -9.2560 |
| S.E_Logq | 0.5081 | 0.5081 | 0.5081 | 0.5081 | 0.5081 |

Terminal year survivor and $F$ summaries:

Age $=1$. Catchability constand w.r.t. time and dependant on age
Year class = 2017
Fleet $=$ UK (E\&W) - BTS -Q3
Survivors $5427.000^{1}$
Raw weights 4.759

```
Age = 2 . Catchability constand w.r.t. time and dependant on age
    Year class = 2016
Fleet = BE-CBT3
Survivors 8450.000
Raw weights 4.218
Fleet = fshk
Survivors 2
Raw weights 0.444
Fleet = UK(E&W)-BTS-Q3
Survivors 13322.000 11480.00
Raw weights 4.436 4.62
\begin{tabular}{lllllll} 
Fleet & Est.Suvivors & Int. s.e. Ext. s.e. Var Ratio N & Scaled Wgts Estimated F \\
{\([1]\),} & "BE-CBT3" & \(" 8450 "\) & \(" 0.48 "\) & "Inf" & "Inf" & "1" "0.307"
\end{tabular}
```

Weighted prediction:
Suvivors Int.s.e. Ext.s.e. Var.Ratio F
[1,] "10752" "" "" "" "0.029"

Age = 3. Catchability constand w.r.t. time and dependant on age
Year class $=2015$
Fleet $=$ BE-CBT3
Survivors
Raw weights $\quad 8.709 \quad 3.588$

Fleet = fshk
Survivors 2767.000
Raw weights 0.444

```
Fleet = UK(E&W)-BTS-Q3
Survivors 13740.000 1981.000 4845.00
Raw weights 2.107 3.773 3.93
\begin{tabular}{|c|c|c|c|c|c|c|c|c|}
\hline & Fleet & Est.Suviv & In & , & Var Ra & N & Scale & \\
\hline [1, ] & "BE-CBT3" & "4305" & "0.263" & "0.028" & "0.106" & "2" & "0.545" & "0.151" \\
\hline [2, ] & "fshk" & "2767" & "1.39" & "Inf" & "Inf" & "1" & "0.02" & "0.226" \\
\hline [3, ] & "UK(E\&W)-BTS-Q3" & "4297" & "0.291" & "0.513" & "1.763" & "3" & "0.435" & "0.151" \\
\hline
\end{tabular}
```

Weighted prediction:

Suvivors Int.s.e. Ext.s.e. Var.Ratio F

```
[1,] "4264" "" "" "" "0.153"
```

Age = 4 . Catchability constand w.r.t. time and dependant on age Year class = 2014

Fleet $=\mathrm{BE}-\mathrm{CBT} 3$
Survivors 4271.0004389 .005737 .000

| Raw weights | 9.003 | 6.93 | 2.778 |
| :--- | :--- | :--- | :--- |

Fleet = fshk
Survivors 2397.000
Raw weights 0.444

Fleet $=\mathrm{UK}(E \& W)-\mathrm{BTS}-\mathrm{Q} 3$
Survivors 3579.0002426 .0002934 .0004417 .000
$\begin{array}{lllll}\text { Raw weights } & 5.069 \quad 1.677 \quad 2.921 \quad 3.043\end{array}$

Fleet Est.Suvivors Int. s.e. Ext. s.e. Var Ratio N Scaled Wgts Estimated F

| [1, ] | "BE-CBT3" | "4508" | "0.199" | "0.072" | "0.362" | "3" | "0.587" | "0.189" |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| [2, ] | "fshk" | "2397" | "1.35" | "Inf" | "Inf" | "1" | "0.014" | "0.33" |

[3,] "UK(E\&W)-BTS-Q3" "3416" "0.237" "0.112" "0.473" "4" "0.399" "0.242"

Weighted prediction:
Suvivors Int.s.e. Ext.s.e. Var.Ratio F
[1,] "4000" "" "" "" "0.21"

Age = 5 . Catchability constand w.r.t. time and dependant on age
Year class $=2013$
Fleet $=\mathrm{BE}-\mathrm{CBT} 3$
Survivors $868.0001113 .000895,000464.000$

| Raw weights | 8.348 | 5.991 | 4.451 | 1.845 |
| :--- | :--- | :--- | :--- | :--- |

Fleet = fshk

|  | 5 |
| :--- | ---: |
| Survivors | $658.000^{2}$ |
| Raw weights | 0.444 |



Age $=7$. Catchability constand w.r.t. time and dependant on age Year class = 2011

| Fleet $=$ | BE-CBT3 |
| :--- | ---: |
|  | 7 |
| Survivors | 830.000 |
| Raw weights | 9.451 |
|  |  |
|  |  |
| Fleet $=$ | fshk |
|  | 7 |
| Survivors | 383.000 |
| Raw weights | 0.444 |

```
Fleet = UK(E&W)-BTS-Q3
Survivors 677.000 903.000 939.000 1477.000 1106.000
Raw weights 1.331 2.345 0.793 1.359 1.416
    Fleet Est.Suvivors Int. s.e. Ext. s.e. Var Ratio N Scaled Wgts Estimated F
[1,] "BE-CBT3" "945" "0.139" "0.072" "0.518" "6" "0.802" "0.161"
[2,] "fshk" "383" "1.383" "Inf" "Inf" "1" "0.011" "0.359"
[3,] "UK(E&W)-BTS-Q3" "982" "0.223" "0.125" "0.561" "5" "0.186" "0.156"
Weighted prediction:
    Suvivors Int.s.e. Ext.s.e. Var.Ratio F
[1,] "942" "" "" "" "0.162"
Age \(=8\). Catchability constand w.r.t. time and dependant on age
    Year class = 2010
Fleet = BE-CBT3
Survivors 208.000 339.000 340.000 338.000 389.000 364.000 166.000
\begin{tabular}{lllllll} 
Raw weights & 4.681 & 6.719 & 4.852 & 3.486 & 2.331 & 1.488
\end{tabular} 0.612
\begin{tabular}{lr} 
Fleet \(=\) fshk \\
Survivors & 243.000 \\
Raw weights & 0.444
\end{tabular}
Fleet = UK(E&W)-BTS-Q3
Survivors 210.000 219.000 250.00 273.000 439.000
Raw weights 0.844 1.312 0.36 0.644 0.671
\begin{tabular}{|c|c|c|c|c|c|c|c|c|}
\hline & leet & t.Suvivors & Int. s. & s & Var Ratio & N & ed Wgts & Estimated \\
\hline [1, ] & "ВЕ-СВТ3" & "308" & "0.142" & "0.093" & "0.655" & "7" & "0.85" & "0.241" \\
\hline [2,] & "fshk" & "243" & "1.326" & "Inf" & "Inf" & "1" & "0.016" & "0.297" \\
\hline [3, ] & "UK(E\&W)-BTS-Q3" & "258" & "0.23" & "0.131" & "0.57" & "5" & "0.135" & "0.282" \\
\hline
\end{tabular}
```

    Weighted prediction:
    Suvivors Int.s.e. Ext.s.e. Var.Ratio F
    [1,] "300" "" "" "" "0.247"

Age $=9$. Catchability constand w.r.t. time and dependant on age Year class = 2009

```
Fleet = BE-CBT3
```

Survivors Raw weights $4.371 \quad 2.934 \quad 3.673 \quad 2.488 \quad 1.712 \quad 1.185$

Fleet = fshk
Survivors 134.000
Raw weights 0.444

```
Fleet = UK(E&W)-BTS-Q3
Survivors 72.000 98.000 22.00 34.00 40.000
Raw weights 0.415 0.667 0.22 0.38
Fleet = UK(E&W)-CBT
Survivors 55.000
Raw weights 0.819
```

    Fleet Est.Suvivors Int. s.e. Ext. s.e. Var Ratio \(N\) Scaled Wgts Estimated F
    [1,] "ВЕ-СВТ3" "70" "0.146" "0.102" "0.699" "8" "0.841" "0.4"
[2,] "fshk" "134" "1.224" "Inf" "Inf" "1" "0.021" "0.229"
[3,] "UK (E\&W)-BTS-Q3" "55" "0.225" "0.262" "1.164" "5" "0.099" "0.486"
[4,] "UK(E\&W)-CBT" "55" "0.331" "Inf" "Inf" "1" "0.039" "0.486"

Weighted prediction:
Suvivors Int.s.e. Ext.s.e. Var.Ratio F
[1,] "68" "" "" "" "0.407"

| Table 36.12-Sol.27.7fg - Fishing mortality |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age/Year | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.0846 | 0.0701 | 0.1077 | 0.0563 | 0.0431 | 0.1327 | 0.0736 | 0.0842 | 0.0729 | 0.2466 |
| 3 | 0.1478 | 0.2590 | 0.3273 | 0.1638 | 0.1261 | 0.4161 | 0.2488 | 0.2216 | 0.1881 | 0.2841 |
| 4 | 0.4004 | 0.2294 | 0.3159 | 0.2182 | 0.1637 | 0.3417 | 0.2729 | 0.2761 | 0.3247 | 0.4417 |
| 5 | 0.4118 | 0.3176 | 0.3238 | 0.2553 | 0.2223 | 0.4416 | 0.2460 | 0.1658 | 0.2437 | 0.4017 |
| 6 | 0.3314 | 0.3520 | 0.2446 | 0.3416 | 0.2717 | 0.2755 | 0.2814 | 0.1991 | 0.2231 | 0.3058 |
| 7 | 0.4256 | 0.2406 | 0.1846 | 0.2329 | 0.2704 | 0.3507 | 0.2924 | 0.1572 | 0.3604 | 0.1176 |
| 8 | 0.3660 | 0.2925 | 0.1687 | 0.2015 | 0.1321 | 0.5246 | 0.2404 | 0.2029 | 0.2697 | 0.2656 |
| 9 | 0.2815 | 0.2246 | 0.2078 | 0.2244 | 0.1869 | 0.3582 | 0.1780 | 0.1953 | 0.2141 | 0.3221 |
| +gp | 0.2815 | 0.2246 | 0.2078 | 0.2244 | 0.1869 | 0.3582 | 0.1780 | 0.1953 | 0.2141 | 0.3221 |
| FBAR 4-8 ${ }^{*}$ | $0.3870^{\prime}$ | $0.2864^{*}$ | $0.2475^{\prime \prime}$ | $0.249{ }^{\prime \prime}$ | $0.2120^{\circ}$ | $0.3868^{\prime \prime}$ | $0.2666^{\prime \prime}$ | $0.200{ }^{\text {² }}$ | $0.2843^{\prime}$ | 0.3065 |
| Age/Year | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.1480 | 0.0857 | 0.1682 | 0.1232 | 0.0503 | 0.1086 | 0.1263 | 0.1140 | 0.1339 | 0.0917 |
| 3 | 0.3824 | 0.2787 | 0.3739 | 0.3069 | 0.3837 | 0.4720 | 0.2819 | 0.2442 | 0.3501 | 0.3984 |
| 4 | 0.3511 | 0.2686 | 0.3739 | 0.3302 | 0.4472 | 0.5296 | 0.5280 | 0.4005 | 0.5040 | 0.6313 |
| 5 | 0.3281 | 0.3202 | 0.3759 | 0.4651 | 0.5181 | 0.5576 | 0.4778 | 0.5526 | 0.4856 | 0.6555 |
| 6 | 0.4353 | 0.3581 | 0.3740 | 0.3948 | 0.4439 | 0.6439 | 0.5594 | 0.6038 | 0.5766 | 0.6988 |
| 7 | 0.3999 | 0.4017 | 0.4827 | 0.5061 | 0.3448 | 0.5144 | 0.8494 | 0.4857 | 0.5889 | 0.6971 |
| 8 | 0.3109 | 0.4051 | 0.7080 | 0.3892 | 0.4606 | 0.5030 | 0.4868 | 0.8251 | 0.5910 | 0.7813 |
| 9 | 0.3995 | 0.4661 | 0.3744 | 0.3259 | 0.4514 | 0.6446 | 0.7071 | 0.6229 | 0.5727 | 0.7830 |
| +gp | 0.3995 | 0.4661 | 0.3744 | 0.3259 | 0.4514 | 0.6446 | 0.7071 | 0.6229 | 0.5727 | 0.7830 |
| FBAR 4-8 ${ }^{\text {² }}$ | $0.3651^{\prime}$ | $0.350{ }^{\text { }}$ | $0.4629^{\prime \prime}$ | $0.4171^{\text {² }}$ | $0.4429^{\prime \prime}$ | $0.5497{ }^{\prime \prime}$ | $0.5803^{\prime \prime}$ | $0.5735^{\prime \prime}$ | $0.5492^{\prime}$ | 0.6928 |
| Age/Year | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.2203 | 0.1283 | 0.0966 | 0.0800 | 0.0446 | 0.0647 | 0.0741 | 0.0429 | 0.1198 | 0.1489 |
| 3 | 0.3073 | 0.3819 | 0.3587 | 0.2880 | 0.4469 | 0.5206 | 0.4641 | 0.3905 | 0.5494 | 0.4189 |
| 4 | 0.4436 | 0.4636 | 0.4061 | 0.5249 | 0.7171 | 0.6699 | 0.5735 | 0.7566 | 0.6407 | 0.3879 |
| 5 | 0.5284 | 0.4836 | 0.4094 | 0.5681 | 0.5674 | 0.5954 | 0.6274 | 0.5521 | 0.6684 | 0.3376 |
| 6 | 0.4904 | 0.4982 | 0.3747 | 0.6477 | 0.6386 | 0.6187 | 0.7634 | 0.5313 | 0.4985 | 0.2530 |
| 7 | 0.4877 | 0.3319 | 0.6053 | 0.5151 | 0.6398 | 0.5003 | 0.7432 | 0.7817 | 0.4510 | 0.3279 |
| 8 | 0.5873 | 0.3047 | 0.5966 | 0.4640 | 0.8185 | 0.5542 | 0.6356 | 0.6718 | 0.4723 | 0.4130 |
| 9 | 0.7420 | 0.5968 | 0.8067 | 0.9195 | 1.0544 | 0.7744 | 1.0370 | 0.5191 | 0.6500 | 0.5253 |
| +gp | 0.7420 | 0.5968 | 0.8067 | 0.9195 | 1.0544 | 0.7744 | 1.0370 | 0.5191 | 0.6500 | 0.5253 |
| FBAR 4-8 ${ }^{\text {² }}$ | $0.5075^{\text { }}$ | $0.416{ }^{\text {² }}$ | 0.4784 | 0.5440 | $0.6763^{\prime \prime}$ | $0.5877^{\text {「 }}$ | $0.6686^{\text {² }}$ | $0.6587^{\prime \prime}$ | $0.5462^{\text {² }}$ | 0.3439 |
| Age/Year | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.1086 | 0.0079 | 0.0230 | 0.1121 | 0.0593 | 0.1748 | 0.1127 | 0.0572 | 0.0736 | 0.0495 |
| 3 | 0.2192 | 0.3081 | 0.2688 | 0.4686 | 0.2911 | 0.3695 | 0.3103 | 0.1961 | 0.1519 | 0.2021 |
| 4 | 0.4044 | 0.3747 | 0.4207 | 0.4929 | 0.3675 | 0.3439 | 0.3232 | 0.3013 | 0.2554 | 0.3433 |
| 5 | 0.4046 | 0.5528 | 0.6390 | 0.5232 | 0.5161 | 0.3783 | 0.3580 | 0.3870 | 0.2747 | 0.3037 |
| 6 | 0.5864 | 0.3802 | 0.8058 | 0.4871 | 0.4531 | 0.2739 | 0.3382 | 0.3261 | 0.3142 | 0.3017 |
| 7 | 0.4201 | 0.4271 | 0.7132 | 0.4967 | 0.3868 | 0.2480 | 0.4089 | 0.3164 | 0.3090 | 0.2602 |
| 8 | 0.3527 | 0.6571 | 0.8040 | 0.3715 | 0.3334 | 0.2859 | 0.3638 | 0.3154 | 0.2404 | 0.3186 |
| 9 | 0.4156 | 0.5684 | 0.8035 | 0.8719 | 0.3957 | 0.4726 | 0.4753 | 0.4540 | 0.2704 | 0.2137 |
| +gp | 0.4156 | 0.5684 | 0.8035 | 0.8719 | 0.3957 | 0.4726 | 0.4753 | 0.4540 | 0.2704 | 0.2137 |
| FBAR 4-8 ${ }^{\text {² }}$ | 0.4336 | $0.478{ }^{\text {² }}$ | $0.6765^{\prime}$ | $0.4743^{\prime \prime}$ | $0.4114^{*}$ | $0.3060^{\text {r }}$ | 0.3584 | $0.329{ }^{\prime \prime}$ | $0.2787^{\prime}$ | 0.3055 |
| Age/Year | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |  |  |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |  |  |
| 2 | 0.0739 | 0.0401 | 0.0828 | 0.0279 | 0.0329 | 0.0662 | 0.0388 | 0.029436 |  |  |
| 3 | 0.2258 | 0.1735 | 0.3574 | 0.1481 | 0.2471 | 0.2060 | 0.1706 | 0.152522 |  |  |
| 4 | 0.3167 | 0.3349 | 0.3682 | 0.4024 | 0.2772 | 0.2656 | 0.3318 | 0.21041 |  |  |
| 5 | 0.2773 | 0.4671 | 0.4423 | 0.3736 | 0.3307 | 0.3164 | 0.1985 | 0.285906 |  |  |
| 6 | 0.3690 | 0.3904 | 0.4284 | 0.3802 | 0.3894 | 0.3257 | 0.2259 | 0.238705 |  |  |
| 7 | 0.3463 | 0.4687 | 0.3990 | 0.4457 | 0.2980 | 0.3927 | 0.2557 | 0.16181 |  |  |
| 8 | 0.3097 | 0.3869 | 0.2948 | 0.4003 | 0.1891 | 0.2907 | 0.3073 | 0.247246 |  |  |
| 9 | 0.2639 | 0.3212 | 0.2920 | 0.5081 | 0.3899 | 0.2127 | 0.1957 | 0.407104 |  |  |
| +gp | 0.2639 | 0.3212 | 0.2920 | 0.5081 | 0.3899 | 0.2127 | 0.1957 | 0.407104 |  |  |
| FBAR 4-8 ${ }^{\text { }}$ | $0.3238{ }^{\text {² }}$ | 0.4096 | $0.3865^{\prime}$ | $0.4004{ }^{\text {² }}$ | $0.2969{ }^{\text { }}$ | $0.3182^{\text {² }}$ | $0.2639{ }^{\prime \prime}$ | 0.2288 |  |  |


| Table 36.13-Sol.27.7fg - Stock numbers at age (start of year, in thousands) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 |
| 1 | 9290 | 4143 | 3291 | 3280 | 2914 | 5142 | 4568 | 5424 | 3490 | 5083 |
| 2 | 5002 | 8406 | 3749 | 2977 | 2968 | 2637 | 4652 | 4133 | 4907 | 3157 |
| 3 | 2065 | 4159 | 7091 | 3045 | 2547 | 2573 | 2090 | 3911 | 3438 | 4128 |
| 4 | 4272 | 1612 | 2905 | 4625 | 2339 | 2031 | 1535 | 1474 | 2836 | 2577 |
| 5 | 1946 | 2590 | 1160 | 1916 | 3364 | 1797 | 1306 | 1057 | 1012 | 1854 |
| 6 | 1613 | 1167 | 1706 | 759 | 1343 | 2437 | 1046 | 924 | 811 | 718 |
| 7 | 1628 | 1048 | 742 | 1209 | 488 | 926 | 1674 | 714 | 685 | 587 |
| 8 | 2617 | 963 | 745 | 559 | 866 | 337 | 590 | 1131 | 552 | 432 |
| 9 | 1611 | 1642 | 650 | 570 | 413 | 687 | 180 | 420 | 835 | 381 |
| +gp | 5212 | 3809 | 3602 | 2822 | 2289 | 1609 | 2348 | 1027 | 921 | 1339 |
| TOTAL | $35258{ }^{\text {² }}$ | 29539 ${ }^{\prime}$ | $25640^{\prime}$ | 21763 | $1953{ }^{\text {² }}$ | $20177^{\text {² }}$ | 19989 ${ }^{*}$ | $20215^{\prime \prime}$ | $19486^{*}$ | 20257 |
|  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| 1 | 4836 | 4856 | 6724 | 4653 | 5581 | 3118 | 5667 | 4437 | 3670 | 8566 |
| 2 | 4599 | 4375 | 4394 | 6084 | 4210 | 5050 | 2821 | 5127 | 4015 | 3320 |
| 3 | 2232 | 3589 | 3634 | 3360 | 4867 | 3623 | 4099 | 2250 | 4139 | 3178 |
| 4 | 2811 | 1378 | 2457 | 2262 | 2237 | 3000 | 2045 | 2798 | 1595 | 2639 |
| 5 | 1499 | 1791 | 953 | 1530 | 1471 | 1294 | 1599 | 1091 | 1696 | 872 |
| 6 | 1123 | 977 | 1176 | 592 | 869 | 793 | 671 | 897 | 568 | 944 |
| 7 | 478 | 657 | 618 | 732 | 361 | 505 | 377 | 347 | 444 | 289 |
| 8 | 472 | 290 | 398 | 345 | 399 | 231 | 273 | 146 | 193 | 223 |
| 9 | 300 | 313 | 175 | 177 | 212 | 228 | 127 | 152 | 58 | 97 |
| +gp | 1208 | 915 | 1074 | 1031 | 693 | 722 | 369 | 604 | 245 | 248 |
| TOTAL | $19559^{\circ}$ | 19142 ${ }^{\prime}$ | $21604{ }^{\text {² }}$ | 20767 | $20900^{*}$ | 18564 | 18047 | 17849 | $16622^{\text {* }}$ | 20376 |
|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 1 | 4177 | 4464 | 4454 | 3427 | 3283 | 3984 | 5450 | 6259 | 14451 | 7909 |
| 2 | 7751 | 3780 | 4039 | 4030 | 3100 | 2970 | 3605 | 4932 | 5664 | 13076 |
| 3 | 2741 | 5627 | 3008 | 3318 | 3366 | 2683 | 2519 | 3029 | 4275 | 4546 |
| 4 | 1930 | 1824 | 3475 | 1902 | 2251 | 1948 | 1442 | 1433 | 1854 | 2233 |
| 5 | 1270 | 1121 | 1038 | 2095 | 1018 | 994 | 902 | 735 | 609 | 884 |
| 6 | 409 | 678 | 625 | 624 | 1074 | 522 | 496 | 436 | 383 | 282 |
| 7 | 425 | 227 | 373 | 389 | 295 | 513 | 255 | 209 | 232 | 211 |
| 8 | 130 | 236 | 147 | 184 | 210 | 141 | 282 | 110 | 87 | 134 |
| 9 | 92 | 65 | 157 | 73 | 105 | 84 | 73 | 135 | 51 | 49 |
| +gp | 323 | 207 | 187 | $253$ | 174 | 205 | $171$ | 225 | 96 | 131 |
| TOTAL | $19249^{\circ}$ | 18228 | 17504 | 16294 | 14876 | $14045^{\text {² }}$ | 15195 ${ }^{\circ}$ | 17502 | 27701 * | 29454 |
|  | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 1 | 4262 | 6766 | 5215 | 5861 | 4958 | 3651 | 4476 | 9941 | 6758 | 2023 |
| 2 | 7157 | 3857 | 6122 | 4718 | 5304 | 4486 | 3304 | 4050 | 8995 | 6115 |
| 3 | 10194 | 5809 | 3462 | 5414 | 3816 | 4523 | 3408 | 2671 | 3461 | 7561 |
| 4 | 2706 | 7408 | 3863 | 2394 | 3066 | 2581 | 2828 | 2261 | 1986 | 2690 |
| 5 | 1371 | 1634 | 4609 | 2295 | 1323 | 1921 | 1656 | 1852 | 1514 | 1392 |
| 6 | 571 | 828 | 851 | 2201 | 1231 | 715 | 1191 | 1047 | 1138 | 1041 |
| 7 | 198 | 287 | 512 | 344 | 1224 | 708 | 492 | 768 | 684 | 752 |
| 8 | 137 | 118 | 170 | 227 | 189 | 752 | 500 | 296 | 507 | 454 |
| 9 | 80 | 87 | 55 | 69 | 142 | 123 | 511 | 314 | 195 | 360 |
| $+\mathrm{gp}$ | 248 | $123$ |  | 99 |  | $142$ | $194$ | $494$ | $854$ | 643 |
| TOTAL | $26924{ }^{\text {² }}$ | $26917{ }^{\text {² }}$ | $25016{ }^{\text {² }}$ | $23622^{*}$ | $21397{ }^{\text {² }}$ | $19601{ }^{\text {² }}$ | 18559 | 23694 | $26091{ }^{\text { }}$ | 23032 |
|  | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |  |  |
| 1 | 4735 | 6372 | 4861 | 3051 | 9333 | 6969 | 13524 | 5997 |  |  |
| 2 | 1830 | 4284 | 5766 | 4398 | 2760 | 8445 | 6306 | 12237 |  |  |
| 3 | 5266 | 1538 | 3724 | 4803 | 3870 | 2417 | 7152 | 5489 |  |  |
| 4 | 5590 | 3802 | 1170 | 2357 | 3748 | 2735 | 1780 | 5456 |  |  |
| 5 | 1727 | 3685 | 2461 | 733 | 1426 | 2570 | 1898 | 1156 |  |  |
| 6 | 930 | 1184 | 2090 | 1431 | 456 | 927 | 1695 | 1408 |  |  |
| 7 | 696 | 582 | 725 | 1232 | 885 | 280 | 606 | 1223 |  |  |
| 8 | 525 | 446 | 329 | 440 | 714 | 595 | 171 | 424 |  |  |
| 9 | 299 | 348 | 274 | 222 | 267 | 535 | 402 | 114 |  |  |
| +gp | 899 | 939 | 687 | 573 | 428 | 512 | 655 | 422 |  |  |
| TOTAL ${ }^{*}$ | $22496{ }^{\text {² }}$ | $23180^{*}$ | $22087{ }^{\text {² }}$ | $19240^{\circ}$ | $23888{ }^{\text {* }}$ | $25984^{*}$ | $34188^{*}$ | 33927 |  |  |


| Table 36.14-Sol.27.7fg - Summary |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | RECRUITS <br> Age 1 | SSB | BIOMASS | LANDINGS | FBAR 4-8 | YIELD/SSB |
| 1971 | 9290 | 7385 | 8809 | 1861 | 0.387 | 0.25 |
| 1972 | 4143 | 5820 | 7427 | 1278 | 0.286 | 0.22 |
| 1973 | 3291 | 4884 | 6175 | 1391 | 0.248 | 0.28 |
| 1974 | 3280 | 5128 | 6113 | 1105 | 0.250 | 0.22 |
| 1975 | 2914 | 4574 | 5403 | 919 | 0.212 | 0.2 |
| 1976 | 5142 | 3958 | 4964 | 1350 | 0.387 | 0.34 |
| 1977 | 4568 | 4322 | 5566 | 961 | 0.267 | 0.22 |
| 1978 | 5424 | 3433 | 4735 | 780 | 0.200 | 0.23 |
| 1979 | 3490 | 3560 | 4754 | 954 | 0.284 | 0.27 |
| 1980 | 5083 | 3714 | 4922 | 1314 | 0.306 | 0.35 |
| 1981 | 4836 | 3191 | 4356 | 1212 | 0.365 | 0.38 |
| 1982 | 4856 | 3301 | 4542 | 1128 | 0.351 | 0.34 |
| 1983 | 6724 | 3439 | 4905 | 1373 | 0.463 | 0.4 |
| 1984 | 4653 | 3695 | 5138 | 1266 | 0.417 | 0.34 |
| 1985 | 5581 | 3146 | 4611 | 1328 | 0.443 | 0.42 |
| 1986 | 3118 | 3200 | 4438 | 1600 | 0.550 | 0.5 |
| 1987 | 5667 | 2395 | 3595 | 1222 | 0.580 | 0.51 |
| 1988 | 4437 | 2556 | 3735 | 1146 | 0.574 | 0.45 |
| 1989 | 3670 | 2000 | 3121 | 992 | 0.549 | 0.5 |
| 1990 | 8566 | 2267 | 3732 | 1189 | 0.693 | 0.52 |
| 1991 | 4177 | 1955 | 3417 | 1107 | 0.507 | 0.57 |
| 1992 | 4464 | 2280 | 3685 | 981 | 0.416 | 0.43 |
| 1993 | 4454 | 2362 | 3721 | 928 | 0.478 | 0.39 |
| 1994 | 3427 | 2142 | 3154 | 1009 | 0.544 | 0.47 |
| 1995 | 3283 | 2048 | 2979 | 1157 | 0.676 | 0.57 |
| 1996 | 3984 | 1996 | 2969 | 995 | 0.588 | 0.5 |
| 1997 | 5450 | 1751 | 2893 | 927 | 0.669 | 0.53 |
| 1998 | 6259 | 1592 | 3013 | 875 | 0.659 | 0.55 |
| 1999 | 14451 | 1792 | 4184 | 1012 | 0.546 | 0.56 |
| 2000 | 7909 | 1926 | 3848 | 1091 | 0.344 | 0.57 |
| 2001 | 4262 | 3073 | 5313 | 1168 | 0.434 | 0.38 |
| 2002 | 6766 | 3938 | 5796 | 1345 | 0.478 | 0.34 |
| 2003 | 5215 | 3568 | 5408 | 1547 | 0.677 | 0.43 |
| 2004 | 5861 | 3119 | 4727 | 1398 | 0.474 | 0.45 |
| 2005 | 4958 | 2995 | 4678 | 1118 | 0.411 | 0.37 |
| 2006 | 3651 | 2536 | 3968 | 946 | 0.306 | 0.37 |
| 2007 | 4476 | 2627 | 3784 | 945 | 0.358 | 0.36 |
| 2008 | 9941 | 2393 | 4289 | 800 | 0.329 | 0.33 |
| 2009 | 6758 | 2758 | 5127 | 805 | 0.279 | 0.29 |
| 2010 | 2023 | 3046 | 4866 | 876 | 0.306 | 0.29 |
| 2011 | 4735 | 3329 | 4814 | 1029 | 0.324 | 0.31 |
| 2012 | 6372 | 3236 | 4613 | 1104 | 0.410 | 0.34 |
| 2013 | 4861 | 2797 | 4430 | 1093 | 0.387 | 0.39 |
| 2014 | 3051 | 2819 | 4102 | 1042 | 0.400 | 0.37 |
| 2015 | 9333 | 2769 | 4586 | 830 | 0.297 | 0.3 |
| 2016 | 6969 | 2867 | 5118 | 831 | 0.318 | 0.29 |
| 2017 | 13524 | 3063 | 5976 | 776 | 0.264 | 0.25 |
| 2018 | 5997 | 3557 | 6404 | 850 | 0.229 | 0.24 |
| Arith. Mean | 5528 | 3131 | 4644 | 1103 | 0.415 | 0.38 |
| Units | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |

## Table 36.15-Sol.27.7fg Input for catch forecast and Fmsy analysis

Input: $\quad$ F in 2019: TAC constraint for 2019 (841 t)
F in 2020-2021: mean 16-18 scaled to 2018
Catch and stock weights: mean 16-18
N age 1 in 2019-2021: GM(71-16)


2020

| Age |  |  | M |  | Mat |  | PF |  | PM |  | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 4975 |  | 0.1 |  | 0 |  | 0 |  | 0 | 0.105333 | 5.87E-17 | 0.121333 |
|  | 2 | 4502 |  | 0.1 |  | 0.14 |  | 0 |  | 0 | 0.148333 | 0.037916 | 0.176667 |
|  | 3 | 4783 |  | 0.1 |  | 0.45 |  | 0 |  | 0 | 0.193667 | 0.149316 | 0.211667 |
|  | 4 | 8773 |  | 0.1 |  | 0.88 |  | 0 |  | 0 | 0.233667 | 0.227949 | 0.261 |
|  | 5 | 3295 |  | 0.1 |  | 0.98 |  | 0 |  | 0 | 0.282 | 0.22598 | 0.306333 |
|  | 6 | 3096 |  | 0.1 |  | 1 |  | 0 |  | 0 | 0.331 | 0.223018 | 0.351 |
|  | 7 | 609 |  | 0.1 |  | 1 |  | 0 |  | 0 | 0.383667 | 0.228627 | 0.418 |
|  | 8 | 775 |  | 0.1 |  | 1 |  | 0 |  | 0 | 0.453 | 0.238504 | 0.474333 |
|  | 9 | 722 |  | 0.1 |  | 1 |  | 0 |  | 0 | 0.48 | 0.23012 | 0.475667 |
| +gp |  | 480 |  | 0.1 |  | 1 |  | 0 |  | 0 | 0.536007 | 0.23012 | 0.55215 |

2021

| Age |  |  | M |  | Mat |  | PF |  | PM |  | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 4975 |  | 0.1 |  | 0 |  | 0 |  | 0 | 0.105333 | 5.87E-17 | 0.121333 |
|  | 2 | 4502 |  | 0.1 |  | 0.14 |  | 0 |  | 0 | 0.148333 | 0.037916 | 0.176667 |
|  | 3 | 3922 |  | 0.1 |  | 0.45 |  | 0 |  | 0 | 0.193667 | 0.149316 | 0.211667 |
|  | 4 | 3728 |  | 0.1 |  | 0.88 |  | 0 |  | 0 | 0.233667 | 0.227949 | 0.261 |
|  | 5 | 6320 |  | 0.1 |  | 0.98 |  | 0 |  | 0 | 0.282 | 0.22598 | 0.306333 |
|  | 6 | 2379 |  | 0.1 |  | 1 |  | 0 |  | 0 | 0.331 | 0.223018 | 0.351 |
|  | 7 | 2241 |  | 0.1 |  | 1 |  | 0 |  | 0 | 0.383667 | 0.228627 | 0.418 |
|  | 8 | 439 |  | 0.1 |  | 1 |  | 0 |  | 0 | 0.453 | 0.238504 | 0.474333 |
|  | 9 | 552 |  | 0.1 |  | 1 |  | 0 |  | 0 | 0.48 | 0.23012 | 0.475667 |
| +gp |  | 864 |  | 0.1 |  | 1 |  | 0 |  | 0 | 0.536007 | 0.23012 | 0.55215 |
| fbar |  |  |  |  |  |  |  |  |  |  |  | 0.228815 |  |

## Table 36.16-Sol.27.7fg - Management option table

F in 2019: TAC constraint for 2019 (841 t)
F in 2020-2021: mean 16-18 scaled to 2018
Catch and stock weights: mean 16-18
N age 1 in 2019-2021: $\mathrm{GM}(71-16)$

2019

| Biomass | SSB | FMult | FBar | Landings |
| :---: | :---: | :---: | :---: | ---: |
| 6924 | 4420 | 0.692005 | 0.15834131 | 766 |


| $\mathbf{2 0 2 0}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| SSB | FMult | FBar | Landings | SSB |
| 5439 | 0.0000 | NA | 0 | 6763 |
| 5439 | 0.1000 | 0.0229 | 137 | 6627 |
| 5439 | 0.2000 | 0.0458 | 272 | 6493 |
| 5439 | 0.3000 | 0.0686 | 404 | 6362 |
| 5439 | 0.4000 | 0.0915 | 533 | 6235 |
| 5439 | 0.5000 | 0.1144 | 659 | 6109 |
| 5439 | 0.6000 | 0.1373 | 783 | 5987 |
| 5439 | 0.7000 | 0.1602 | 904 | 5867 |
| 5439 | 0.8000 | 0.1831 | 1022 | 5750 |
| 5439 | 0.9000 | 0.2059 | 1138 | 5635 |
| 5439 | 1.0000 | 0.2288 | 1252 | 5523 |
| 5439 | 1.1000 | 0.2517 | 1363 | 5413 |
| 5439 | 1.2000 | 0.2746 | 1472 | 5305 |
| 5439 | 1.3000 | 0.2975 | 1579 | 5200 |
| 5439 | 1.4000 | 0.3203 | 1684 | 5097 |
| 5439 | 1.5000 | 0.3432 | 1786 | 4997 |
| 5439 | 1.6000 | 0.3661 | 1886 | 4898 |
| 5439 | 1.7000 | 0.3890 | 1984 | 4802 |
| 5439 | 1.8000 | 0.4119 | 2080 | 4707 |
| 5439 | 1.9000 | 0.4348 | 2174 | 4615 |
| 5439 | 2.0000 | 0.4576 | 2267 | 4524 |

Input units are thousands and kg - output in tonnes

| 2020 |  |  | 2021 | 2021-2020 | 2020-2019 |  |
| :---: | ---: | ---: | ---: | ---: | ---: | :---: |
| FMult | Landings | FBar | SSB | SSB change | TAC change | Basis |
| 1.298 | 1577 | 0.297 | 5202 | $-4.4 \%$ | $105.8 \%$ | msyapproach |
| 1.298 | 1577 | 0.297 | 5202 | $-4.4 \%$ | $105.8 \%$ | Fmsy |
| 2.1808 | 2428 | 0.499 | 4366 | $-19.7 \%$ | $216.9 \%$ | Fmsy_upper |
| 0.7211 | 929 | 0.165 | 5842 | $7.4 \%$ | $21.2 \%$ | Fmsy_lower |
| 0.6568 | 766 | 0.13417 | 6003 | $10.4 \%$ | $0.0 \%$ | TACstable |
| 0.7637 | 881 | 0.15581 | 5890 | $8.3 \%$ | $15.0 \%$ | TACplus15 |
| 0.5523 | 651 | 0.11295 | 6117 | $12.5 \%$ | $-15.0 \%$ | TACminus15 |
| 1.8355 | 2114 | 0.42 | 4674 | $-14.1 \%$ | $175.9 \%$ | Fpa |
| 2.5261 | 2721 | 0.578 | 4080 | $-25.0 \%$ | $255.1 \%$ | Flim |
| 0.7651 | 894 | 0.15834 | 5877 | $8.1 \%$ | $16.7 \%$ | FInt |
| 5.6814 | 4644 | 1.31261 | 2228 | $-59.0 \%$ | $506.1 \%$ | Btrigger |
| 7.6075 | 5325 | 1.753 | 1592 | $-70.7 \%$ | $595.0 \%$ | Blim |


| Table 36.17 - Sol.27.7fg - Detailed <br> F in 2019: TAC constraint for 2019 ( 841 t) <br> F in 2020-2021: mean 16-18 scaled to 2018 <br> Catch and stock weights: mean 16-18 <br> $N$ age 1 in 2019-2021: GM(71-16) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
| Year: Age | $2019 \text { F }$ | F multiplier: CatchNos | $\begin{gathered} 0.76511 \\ \quad \text { Yield } \end{gathered}$ | Fbar: <br> StockNos |  | $0.15834$ <br> Biomass | SSNos | SSB |
| 1 | 0.00000 | 0 | 0 | 4975 |  | 524 | 0 | 0 |
| 2 | 0.02624 | 134 | 24 | 5427 |  | 805 | 760 | 113 |
| 3 | 0.10333 | 1005 | 213 | 10751 |  | 2082 | 4838 | 937 |
| 4 | 0.15774 | 593 | 155 | 4264 |  | 996 | 3752 | 877 |
| 5 | 0.15638 | 552 | 169 | 4000 |  | 1128 | 3920 | 1105 |
| 6 | 0.15433 | 107 | 38 | 786 |  | 260 | 786 | 260 |
| 7 | 0.15821 | 140 | 58 | 1003 |  | 385 | 1003 | 385 |
| 8 | 0.16505 | 137 | 65 | 942 |  | 427 | 942 | 427 |
| 9 | 0.15924 | 42 | 20 | 300 |  | 144 | 300 | 144 |
| 10 | 0.15924 | 45 | 25 | 322 |  | 173 | 322 | 173 |
| Total | 0.15834 | 2755 | 766 | 32770 |  | 6924 | 16623 | 4420 |
| Year: Age | $2020$ $F$ | F multiplier: <br> CatchNos | $\begin{aligned} & 1 \\ & \text { Yield } \end{aligned}$ | Fbar: <br> StockNos | - | $0.22882$ <br> Biomass | SSNos | SSB |
| 1 | 0.000 | 0 | 0 | 4975 |  | 524 | 0 | 0 |
| 2 | 0.038 | 159 | 28 | 4502 |  | 668 | 630 | 93 |
| 3 | 0.149 | 632 | 134 | 4783 |  | 926 | 2152 | 417 |
| 4 | 0.228 | 1705 | 445 | 8773 |  | 2050 | 7721 | 1804 |
| 5 | 0.226 | 635 | 195 | 3295 |  | 929 | 3229 | 911 |
| 6 | 0.223 | 590 | 207 | 3096 |  | 1025 | 3096 | 1025 |
| 7 | 0.229 | 119 | 50 | 609 |  | 234 | 609 | 234 |
| 8 | 0.239 | 157 | 74 | 775 |  | 351 | 775 | 351 |
| 9 | 0.230 | 142 | 67 | 722 |  | 347 | 722 | 347 |
| 10 | 0.230 | 94 | 52 | 480 |  | 257 | 480 | 257 |
| Total | 0.22882 | 4233 | 1252 | 32010 |  | 7311 | 19414 | 5439 |
| Year: Age | $2021 \text { F }$ | F multiplier: CatchNos | 1 <br> Yield | Fbar: <br> StockNos | F | 0.22882 <br> Biomass | SSNos | SSB |
| 1 | 0.000 | 0 | 0 | 4975 |  | 524 | 0 | 0 |
| 2 | 0.038 | 159 | 28 | 4502 |  | 668 | 630 | 93 |
| 3 | 0.149 | 518 | 110 | 3922 |  | 759 | 1765 | 342 |
| 4 | 0.228 | 724 | 189 | 3728 |  | 871 | 3280 | 766 |
| 5 | 0.226 | 1219 | 373 | 6320 |  | 1782 | 6194 | 1747 |
| 6 | 0.223 | 453 | 159 | 2379 |  | 787 | 2379 | 787 |
| 7 | 0.229 | 437 | 183 | 2241 |  | 860 | 2241 | 860 |
| 8 | 0.239 | 89 | 42 | 439 |  | 199 | 439 | 199 |
| 9 | 0.230 | 108 | 52 | 552 |  | 265 | 552 | 265 |
| 10 | 0.230 | 169 | 94 | 864 |  | 463 | 864 | 463 |
| Total | 0.22882 | 3876 | 1229 | 29922 |  | 7179 | 18344 | 5523 |

Figure 36.1 - Sol. 27.7 fg - Dotted lines give the length distributions of UK (England and
Wales) landings; solid lines of Belgian landings


Figure 36.2 - Sol.27.7fg - Age composition of landings


Figure 36.3-Sol.27.7fg - Standardized catch proportion
Standardized catch (L) proportion at age


Figure 36.4a-Sol.27.7fg - Belgian length distributions of discarded and retained fish from discard sampling studies


Figure 36.4b-Sol.27.7fg - UK (E+W) Length distributions of discarded and retained fish from discard sampling studies


Figure 36.4c - Sol.27.7fg -
Ireland Length distributions of discarded and retained fish from discard sampling studies



Figure 36.5a-Sol.27.7.fg - Mean-standardised indices



Figure 36.5b. Sol.27.7.fg - Mean-standardised indices.

Figure 36.6 - Sol.27.7fg - Consistency plot UK(E\&W)-BTS-Q3 survey


Figure 36.7 - Sol.27.7.fg - Effort (hours ('000) (BE-CBT and IR-CBT), hours ('000) GRT corrected (UK-CBT)) and LPUE (kg/hour (BE-CBT and IR-CBT), kg/hour GRT corrected (UK-


Figure 36.8 - Sol.27.7fg - Consistency plot UK(E\&W) commercial beam trawl


Figure 36.9a - Sol.27.7fg -Consistency plot commercial Belgian beam trawl (1971-1996)


Figure 36.9b - Sol.27.7fg - Consistency plot commercial Belgian beam trawl (2006-2018)


Figure $\mathbf{3 6 . 1 0}$ - Sol.27.7fg - Catchability residuals for the final XSA run


Figure 36.11-Sol.27.7.fg - Estimates of survivors from different fleets and shrinkage, as well as their different weighting in the final XSA-run



Figure 36.12 - Sol.27.7fg - Retrospective XSA analysis (shinkage SE=1.5)


Figure 36.13 - Sol.27.7fg - Summary plots


Figure $\mathbf{3 6} .14$ - Sol.27.7fg - Comparison with last year's assessment


Figure 36.15-Sol.27.7fg -
Year-class sources and contributions for the short-term forecast


Figure $\mathbf{3 6 . 1 6}$ - Sol.27.7fg - Three year average exploitation pattern, standardised to Fbar (4-8)


Figure 36.17 - Sol.27.7fg - Stock/recruitment plot


## 35 Sole (Solea solea) in divisions 7.h-k (Celtic Sea South, southwest of Ireland)

## Type of assessment in 2019

An update XSA assessment was performed for the $7 . \mathrm{jk}$ component of the landings according to the stock annex. Only MSY reference points were explored as they are comparable with the XSA.

## ICES advice applicable to 2019

ICES advises that when the precautionary approach is applied, catches in 2019 should be no more than 213 tonnes.
http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2019/2019/sol-7h-k.pdf

## ICES advice applicable to 2018

Based on ICES approach to data-limited stocks, ICES advises that catches should be no more than 268 t in 2018. All catches are assumed to be landed.
http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2017/2017/sol.27.7h-k.pdf

### 35.1 General

## Stock description and management units

Sole in 7.j are mainly caught by Irish vessels on sandy grounds off the southwest of Ireland. Catches in 7.k are negligible. 7.h is also considered part of the stock for assessment purposes, but there is no evidence to suggest that this is actually the same stock (Figure 33.1). Irish VMS and logbook data indicate that the 7.j landings occur close to shore and this species is a small (but valuable) component (up to 5\%) of the landings in a mixed fishery.

The TAC is set for divisions $7 . h, j$ and $k$. However, as historically no age-disaggregated data were available for $7 . \mathrm{h}$, the assessment is performed for 7.jk only.

Management applicable to 2019 and 2018

TAC table 2019

| Species: | Common sole Solea solea |  | Zone: | $7 \mathrm{~h}, 7 \mathrm{j}$ and 7 k <br> (SOL/7H)K.) |
| :---: | :---: | :---: | :---: | :---: |
| Belgium |  | 32 |  |  |
| France |  | 64 |  |  |
| Ireland |  | 171 |  |  |
| The Netherlands |  | 51 |  |  |
| United Kingdom |  | 64 |  |  |
| Union |  | 382 |  |  |
| TAC |  | 382 |  | Precautionary TAC <br> Article 7(2) of this Regulation applies <br> Article 13(1) of this Regulation applies |

TAC table 2018


Article 12(1) refers to the closure of the Porcupine Bank in May and July.

## Landings obligation

In 2016 the landings obligation will apply to this stock for the first time. According to the regulation (EC, 2015) vessels where more than $5 \%$ of their landings using beam trawls were sole during the reference years (2013 and 2014) in ICES divisions 7.b, 7.c and 7.f-7k will be covered by the Landings Obligation. The landings obligation will also apply to all catches of sole with trammelnets or gillnets. These vessels will have to land all sole in 2016. However, a de minimis exemption will also apply allowing for up to a maximum of $3 \%$ of the annual catch to be discarded. Given the low discards observed in the fishery the landings obligation is unlikely to have a significant impact on this stock or the advice given.

### 35.2 Data

### 35.2.1 Landings and discards

The nominal landings are given in Table 33.1. Historic Belgian landings from 7.j are considered to have been area misreported and have been removed from the total landings. Because age data were only available for Irish landings (which were mainly from 7.jk) the remainder of Section 37 concerns 7.jk only.

Table 33.2 gives the landings in 7.jk. Generally, Ireland has taken around $90 \%$ of the landings.

Discarding of sole in $7 . \mathrm{jk}$ is not considered to be a problem. Only three of the 13 observer trips noted discards of sole (Figure 33.2).

### 35.2.2 Landings numbers-at-age

Landings numbers-at-age are given in Table 33.2 and Figure 33.3. Figure 33.4 shows a bubbleplot of the standardised landings proportions-at-age. The numbers-at-age matrix shows quite good cohort tracking, suggesting that ageing is accurate and that recruitment is variable. Figure 33.5 gives the stock weights (which are the same as the landings weights).

### 35.2.3 Biological

Natural mortality was assumed to be 0.1 for all ages and the proportion mature is assumed to be as follows:

| Age 2 Age 3 | Age 4 | Age 5 | Age 6+ |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.14 | 0.45 | 0.88 | 0.98 | 1.00 |

### 35.2.4 Surveys and commercial tuning fleets

Historically, no survey index was available for this stock as the Irish IBTS Q4 ground fish Survey data were considered too noisy to be used. The development of the Irish Beam Trawl Ecosystem Survey (IBES) may now represent a suitable index. The first of these surveys took place in 2016 (ICES, 2016c) and was repeated in 2017 and 2018. The addition of this index would need to be considered at a benchmark. A commercial tuning index is available which uses Irish VMS data linked to logbook landings (see Gerritsen et al., 2011 for details on linking VMS and logbook data). The data were used to identify an area where sole are caught by OTB vessels (Figure 33.6). Next, the effort and landings of the OTB vessels inside the sole area were estimated. The VMSbased lpue showed similar trends to the lpue of Irish OTB vessels in the whole of 7.j, however by limiting the spatial extent, the index will be less sensitive to changes in the spatial distribution of the fleet. All vessels operating in this area are assumed to be capable of catching sole, however this is not the case, as offshore vessels catch little to no sole.

The age composition of the Irish OTB fleet in 7.j was used for the tuning fleet (Table 33.4). Figure 33.7 shows the log-standardised numbers-at-age in the tuning index by year and cohort. No year effects are obvious, and cohort tracking appears to be reasonably good, however, there is a clear decrease in the age 3 in 2018. Figure 33.8 shows the internal consistency regressions for the tuning fleet.

### 35.2.5 Data quality

Sampling appears to be sufficient to establish catch numbers-at-age. The tuning index is quite short and does, but should be long enough to inform the trends that are not already converged.

### 35.3 Historical stock assessment development

Target category: 3.2.0.
Model used: XSA.

Software used: Lowestoft vpa95.exe and FLR with R version 3.5.1 and packages FLXSA_2.6.2, FLAssess_2.6.3, FLEDA_2.5.2, FLCore_2.6.13.

### 35.3.1 Exploratory assessment

Several exploratory assessments were carried out by means of a separable VPA and XSA. The initial VPA runs explored the year and age range to be used in the separable and the choices of reference age, final F and S. The XSA runs explored the choices of q-age, F-shrinkage and the minimum SE threshold. The results of these are available on the ICES SharePoint site of WGCSE under data for this stock.

### 35.3.2 Final assessment

The model was applied to catch numbers for ages $2-10+$ for the years 1993-2018. The tuning fleet included ages 3-9 for the years 2006-2018.

Model Options:

| Option | Setting |
| :--- | :---: |
| Ages catch dep stock size | None |
| Q plateau | No |
| Taper | 7 |
| F shrinkage SE | 1.5 |
| F shrinkage year range | 5 |
| F shrinkage age range | No |
| Fleet SE threshold | 5 |
| Prior weights | 0.2 |

The diagnostics of the final XSA assessment are given in Table 33.5. Figure 33.9 shows the residuals. There are some year effects but the absolute values are small. Because the catch and the tuning fleet have nearly identical age compositions, the year effects result from the lpue estimate of the tuning fleet.

A Mohn's rho analysis was conducted based on the XSA stock assessment results, i.e. the last data year (2018) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (ages 3-9) | recruitment |
| :--- | :--- | :--- | :--- |
| Mohn's rho value | -0.00759231890293091 | 0.0406781290651195 | 0.00784258485351308 |

The Mohn's rho values for this assessment are very low and well below the threshold of $20 \%$ imposed by ICES for 2019 assessments, i.e. the current assessment indicates a high consistency.

### 35.3.3 State of the stock

The summary table with a time-series of landings, recruitment, SSB and F is given in Table 33.6 and Figure 33.10. Recruitment is variable without a clear trend. The SSB has declined from nearly 800 tonnes to around 400 t in 2000-2009 but appears to have recovered to around 800 t in recent years F shows a slowly declining trend and currently appears to be quite low, with another revision down in 2018.

### 35.4 MSY evaluation

Previously for this stock WKProxy (ICES, 2016a) proposed an Fmsy reference point of $\mathrm{F}=0.17$, based on $\mathrm{F}_{0.1}$ from a Thompson-Bell yield-per-recruit analysis of the landings numbers-at-age. This is a data-limited approach (which was in line with the ToRs of WKProxy); however, the resulting reference point is not directly comparable with the outputs from the XSA (only the landings data are used in the Thompson-Bell approach). In 2016 this working group (ICES, 2016d) recommended that it would be more appropriate to move the stock to Category 2 next year and to apply the WKMSYREF4 (ICES, 2016b) methodology for estimating reference points (ICES, 2012).

An exploratory MSY evaluation was completed by WGCSE in 2017 (ICES, 2017), which followed the WKMSYREF4 guidelines. These reference points were not revaluated in 2018, but were applied as estimated in 2017 (ICES, 2017). As there is no obvious stock-recruitment relationship, it is difficult to specify an appropriate SR model. The SR estimation was carried out on age $>=3$ as that is the onset of recruitment using: fit <-eqsr_fit_shift (stock, nsamp $=1000$, mode/s $=$ $c$ ("Segreg"), rshift $=3$ ). From this Blim was estimated to be 424.88 (Blim <- median(fit\$sr.sto\$b.b)) and a $B_{p a}$ at 590.41 ( $B_{p a}<-B_{p a}\left(B_{l i m}, 0.2\right)$. The following settings were used to estimate the MSY reference points using the eqsim_run\{msy\} function in the MSY package in R (full code available on SharePoint):

$$
\begin{aligned}
& \text { stocksetup <- list }(\text { data }=\text { stock, } \\
& \text { bio.years }=c(2007,2016), \\
& \text { bio.const }=\text { FALSE, } \\
& \text { sel. years }=c(2007,2016), \\
& \text { sel.const }=\text { FALSE, } \\
& \text { Fscan }=\text { seq }(0,0.44, b y=0.005), \\
& \text { Fcv }=0.212, \\
& \text { Fphi }=0.423, \\
& \text { Blim }=\text { Blim, } \\
& \text { Bpa }=\text { Bpa, } \\
& \text { verbose }=\text { TRUE, } \\
& \text { extreme.trim }=c(0.05,0.95)
\end{aligned}
$$

Where $\mathrm{F}_{\mathrm{cv}}$ and $\mathrm{F}_{\text {phi }}$ were the same as those used by WKMSYREF4 for plaice in 7.e (ICES, 2016b), which was calculate during WKMSYREF3 (ICES, 2014). Figures 33.12 and 33.13 summarise the MSY evaluation. The analysis resulted in an estimate of $\mathrm{F}_{\mathrm{MSY}}=0.161$ without a $\mathrm{B}_{\text {trigger }}$ harvest control rule and $F_{M S Y}=0.181$ with a $B_{\text {trigger }}=B_{\text {PA }} H C R$. These values are slightly higher than the Fmsy proxy of 0.25 proposed by WKProxy (ICES, 2016a).

### 35.4.1 MSY and Biological reference points

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 590 | $\mathrm{B}_{\mathrm{pa}}$ | ICES (2017) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.161 | Median point estimates of EqSim with segmented regression S-R relationship | ICES (2017) |
| Precautionary approach | $\mathrm{Bl}_{\text {lim }}$ | 425 | Breakpoint segmented regression S-R relationship | ICES (2017) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 590 | $\mathrm{Bl}_{\lim } \times \exp (1.645 \times \sigma) ; \sigma=0.20$ | ICES (2017) |
|  | $\mathrm{F}_{\text {lim }}$ | 0.222 | F with $50 \%$ probability of $\mathrm{SSB}<\mathrm{Bl}_{\mathrm{lim}}$ | ICES (2017) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.161 | $\mathrm{F}_{\text {lim }} \times \exp (-1.645 \times \sigma) ; \sigma=0.20$ | ICES (2017) |
| Management plan | SSB ${ }_{\text {mgt }}$ |  |  |  |
|  | $F_{\text {mgt }}$ |  |  |  |

### 35.5 Uncertainties and bias in the assessment and forecast

The assessment is carried out on the 7.jk part of the stock area only.
There is sufficient contrast in the landings-at-age matrix to inform the model. However, there may be some data issues between 1999 and 2003, which result in erratic F estimates.

The use of a commercial tuning fleet has the potential to introduce bias if the behaviour or efficiency of the fleet changes. E.g. changes to the gear, vessel power, towing speed, etc. can influence the catch rates. By limiting the index to an area where sole is known to be caught, some of the potential bias due to changes in spatial effort distribution will be avoided. The working group applied a spatial stratification to check that changes in effort distribution within the sole area did not affect the index and this did not appear to be the case. Because the stratified estimate is likely to be less precise, the final tuning index was based on the un-stratified estimate. More sophisticated modelling approaches to standardise the commercial index could be investigated for a future benchmark.

### 35.6 Recommendations for the next benchmark

This stock is scheduled to be benchmarked in 2020.

### 35.7 Management considerations

Fishing mortality has been slowly declining in the last ten years and SSB has been stable in recent years.

The TAC area includes Division 7h. However, the landings from divisions 7jk are taken in the north eastern part of Division 7j which is remote from the northern part of Division 7h, where most of the Division 7h landings are taken. It is likely that the sole from Division 7h are part of the divisions 7e or 7fg stocks. No further information on stock structure is likely to become available in the short term.

The catches are taken in a mixed fisheries and should be managed as such. Constraining the landings by TAC will not constrain the catches. Because sole are caught in spatially distinct areas, restricting effort in these areas will be more effective than limiting landings. The catches are taken in a mixed fisheries and should be managed as such. Constraining the landings by TAC will not constrain the catches. The TAC is currently not restrictive, but for some countries, the quota appears to have become restrictive.

### 35.8 References

Gerritsen HD and Lordan C. 2011. Integrating Vessel Monitoring Systems (VMS) data with daily catch data from logbooks to explore the spatial distribution of catch and effort at high resolution. ICES J Mar Sci 68 (1): 245-252.

ICES. 2012. ICES implementation of advice for data limited stocks in 2012. Report in support of ICES advice. ICES CM 2012/ACOM:68.

ICES. 2014. Report of the Joint ICES-MYFISH Workshop to consider the basis for FMSY ranges for all stocks (WKMSYREF3), 17-21 November 2014, Charlottenlund, Denmark. ICES CM 2014/ACOM:64. 164 pp.

ICES. 2016a. Report of the Workshop to consider MSY proxies for stocks in ICES category 3 and 4 stocks in Western Waters (WKProxy), 3-6 November 2015, ICES Headquarters, Copenhagen. ICES CM 2015/ACOM:61. 183 pp.

ICES. 2016b. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

ICES. 2016c. Final Report of the Working Group on Beam Trawl Surveys (WGBEAM), 12-15 April 2016, La Rochelle, France. ICES CM 2016/ SSGIEOM:20. 125 pp.

ICES. 2016d. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 4-13 May 2016, Copenhagen, Denmark. ICES CM 2016ACOM:13. 1312 pp.

Table 39.1. Sole in Divisions 7.h-k (Southwest Ireland). Nominal landings ( $\mathbf{t}$ ), 1993-2018, as officially reported to ICES. Belgian landings from 7.j are considered to have been area-misreported and are not included in the total. * Preliminary data.

| Row Labels | $\begin{gathered} 7 \mathrm{~h} \\ \text { BEL } \end{gathered}$ | FRA | IRE | NL | UK | $\begin{gathered} 7 \mathbf{j} \\ \text { BEL } \end{gathered}$ | FRA | IRE | UK | $\begin{aligned} & 7 \mathrm{k} \\ & \text { FRA } \end{aligned}$ | IRE | UK | 7 h Total TOT | 7 jk Total TOT | 7 hjk <br> TOT | 7hjk <br> WG Est |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1993 |  | 43 |  |  | 206 |  | 1 | 237 | 8 |  |  |  | 249 | 246 | 495 |  |
| 1994 |  | 42 | 8 |  | 172 |  |  | 176 | 2 |  |  |  | 222 | 178 | 400 |  |
| 1995 |  | 44 | 11 |  | 186 |  | 1 | 232 | 6 | 2 |  |  | 241 | 241 | 482 |  |
| 1996 |  | 48 | 20 | 70 | 147 |  | 2 | 162 | 1 |  | 1 |  | 285 | 166 | 451 | 443 |
| 1997 |  | 56 | 16 |  | 111 |  | 2 | 187 | 1 |  |  | 1 | 183 | 191 | 374 | 564 |
| 1998 |  | 65 | 13 | 7 | 109 |  | 8 | 208 | 2 | 1 |  |  | 194 | 219 | 413 | 423 |
| 1999 | 5 |  | 8 | 1 | 96 | 96 |  | 199 | 1 |  |  |  | 110 | 200 | 310 | 381 |
| 2000 |  | 72 | 8 | 10 | 95 | 8 | 4 | 103 |  | 2 |  |  | 185 | 109 | 294 | 329 |
| 2001 | 6 | 86 | 11 |  | 111 | 7 | 11 | 113 |  | 2 | 1 |  | 214 | 127 | 341 | 325 |
| 2002 | 85 | 85 | 9 |  | 124 | 69 | 8 | 120 |  | 15 | 1 |  | 303 | 144 | 447 | 430 |
| 2003 | 122 | 113 | 23 |  | 78 | 48 | 20 | 82 |  |  |  |  | 336 | 102 | 438 | 245 |
| 2004 | 155 | 95 | 33 |  | 79 | 2 | 7 | 78 |  |  |  |  | 362 | 85 | 447 | 290 |
| 2005 | 90 | 86 | 28 |  | 112 |  | 7 | 69 |  |  | 1 |  | 316 | 77 | 393 | 326 |
| 2006 | 36 | 81 | 14 | 1 | 86 | 0 | 11 | 49 | 1 | 0 | 0 | 0 | 218 | 61 | 279 | 272 |
| 2007 | 31 | 69 | 4 | 0 | 91 | 0 | 9 | 73 | 0 | 0 | 1 | 0 | 195 | 83 | 278 | 277 |
| 2008 | 10 | 49 | 3 | 0 | 80 | 0 | 8 | 69 | 0 | 0 | 0 | 0 | 142 | 77 | 219 | 225 |


| Row Labels | 7 h |  |  |  |  | 7 j |  |  |  | 7 k |  |  | 7 h Total | 7 jk Total | 7 hjk <br> TOT | 7hjk <br> WG Est |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | BEL | FRA | IRE | NL | UK | BEL | FRA | IRE | UK | FRA | IRE | UK | TOT | TOT |  |  |
| 2009 | 11 | 70 | 0 | 0 | 58 | 0 | 9 | 60 | 0 | 0 | 0 | 0 | 139 | 69 | 208 | 208 |
| 2010 | 20 | 73 | 3 | 0 | 51 | 0 | 14 | 68 | 0 | 0 | 0 | 0 | 147 | 82 | 229 | 228 |
| 2011 | 10 | 70 | 1 | 0 | 54 | 0 | 23 | 63 | 0 | 1 | 0 | 0 | 135 | 87 | 222 | 237 |
| 2012 | 18 | 74 | 2 | 0 | 46 | 0 | 11 | 83 | 0 | 0 | 0 | 0 | 140 | 94 | 234 | 228 |
| 2013 | 4 | 69 | 1 | 0 | 47 | 0 | 7 | 84 | 0 | 0 | 0 | 0 | 121 | 91 | 212 | 211 |
| 2014 | 42 | 56 | 3 | 0 | 54 | 0 | 5 | 82 | 0 | 0 | 0 | 0 | 155 | 87 | 242 | 243 |
| 2015 | 40 | 70 | 3 | 0 | 53 | 0 | 4 | 74 | 0 | 0 | 0 | 0 | 166 | 78 | 244 | 248 |
| 2016 | 91 | 67 | 4 | 0 | 61 | 0 | 10 | 94 | 2 | 0 | 0 | 0 | 223 | 115 | 329 | 344 |
| 2017 | 70 | 75 | 4 | 0 | 38 | 4 | 6 | 81 | 1 | 1 | 1 | 0 | 188 | 92 | 280 | 295 |
| 2018* | 92 | 84 | 8 | 0 | 33 | 5 | 7 | 53 | <1 | <1 | <1 | 0 | 217 | 66 | 282 | 294 |

Table 33.2. Landings numbers-at-age for sole in 7.jk.

|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 32.8 | 217.9 | 224.5 | 76.8 | 55.7 | 56.7 | 31.5 | 20.6 | 11.6 | 11.0 | 5.5 | 4.7 | 4.7 | 8.2 | 0.9 |
| 1994 | 23.5 | 117.3 | 130.2 | 68.8 | 40.8 | 22.4 | 19.1 | 10.9 | 12.0 | 13.0 | 10.7 | 4.0 | 3.3 | 11.0 | 12.1 |
| 1995 | 0.0 | 279.2 | 80.8 | 174.0 | 117.1 | 50.9 | 14.9 | 15.3 | 4.1 | 22.0 | 7.7 | 8.5 | 2.1 | 2.2 | 2.1 |
| 1996 | 12.3 | 45.9 | 115.9 | 80.4 | 52.7 | 54.2 | 31.5 | 8.1 | 4.8 | 5.6 | 10.0 | 2.6 | 5.3 | 6.3 | 20.9 |
| 1997 | 39.0 | 160.9 | 83.5 | 109.7 | 42.6 | 41.5 | 37.7 | 15.7 | 1.4 | 0.0 | 3.9 | 3.0 | 3.2 | 2.2 | 11.4 |
| 1998 | 23.5 | 137.2 | 113.3 | 58.9 | 92.7 | 40.0 | 43.1 | 34.4 | 8.8 | 5.4 | 2.8 | 5.0 | 2.8 | 0.0 | 29.7 |
| 1999 | 51.2 | 179.3 | 217.7 | 187.0 | 67.0 | 76.9 | 30.1 | 27.6 | 19.1 | 1.8 | 10.5 | 1.3 | 0.0 | 1.2 | 18.1 |
| 2000 | 39.4 | 95.6 | 82.9 | 41.8 | 28.9 | 15.8 | 20.9 | 10.8 | 16.6 | 7.9 | 3.0 | 0.0 | 1.7 | 0.0 | 3.4 |
| 2001 | 64.6 | 114.6 | 52.6 | 49.1 | 37.7 | 22.4 | 21.8 | 14.3 | 9.2 | 3.6 | 2.0 | 5.2 | 3.0 | 1.7 | 3.3 |
| 2002 | 12.7 | 139.3 | 183.5 | 65.7 | 37.9 | 38.7 | 15.0 | 8.3 | 24.1 | 7.7 | 20.6 | 5.2 | 5.5 | 3.2 | 22.1 |
| 2003 | 2.0 | 53.6 | 92.6 | 128.0 | 76.2 | 44.9 | 18.4 | 3.9 | 5.4 | 8.9 | 13.7 | 0.0 | 2.7 | 0.9 | 5.4 |
| 2004 | 7.0 | 18.4 | 92.3 | 47.8 | 36.4 | 18.7 | 13.7 | 5.9 | 8.0 | 1.2 | 6.9 | 1.2 | 4.5 | 3.4 | 12.3 |
| 2005 | 9.4 | 34.0 | 47.2 | 64.7 | 17.1 | 38.3 | 20.7 | 9.4 | 3.8 | 4.2 | 0.0 | 3.8 | 4.4 | 3.2 | 6.7 |
| 2006 | 12.8 | 29.1 | 29.7 | 27.6 | 37.7 | 17.8 | 15.7 | 10.8 | 6.0 | 3.8 | 1.3 | 0.6 | 1.4 | 1.3 | 8.6 |
| 2007 | 1.1 | 44.0 | 35.7 | 30.1 | 44.4 | 42.3 | 20.5 | 15.9 | 10.1 | 4.3 | 4.2 | 1.2 | 3.3 | 1.1 | 3.3 |
| 2008 | 1.2 | 24.7 | 89.6 | 42.6 | 21.5 | 20.3 | 25.0 | 10.5 | 7.9 | 4.8 | 2.8 | 3.2 | 2.0 | 1.4 | 3.9 |
| 2009 | 0.3 | 14.8 | 38.4 | 76.5 | 31.4 | 16.9 | 16.6 | 15.9 | 6.3 | 6.1 | 5.5 | 1.0 | 0.8 | 0.0 | 3.2 |


|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | 5.0 | 48.5 | 49.5 | 54.0 | 47.3 | 13.7 | 8.8 | 9.1 | 8.8 | 6.2 | 6.7 | 2.9 | 3.1 | 0.2 | 4.8 |
| 2011 | 0.7 | 24.3 | 65.1 | 46.3 | 32.8 | 32.7 | 13.5 | 8.4 | 8.4 | 7.6 | 6.9 | 4.4 | 2.2 | 1.0 | 8.4 |
| 2012 | 0.7 | 11.4 | 48.1 | 70.8 | 33.6 | 31.0 | 26.4 | 9.8 | 9.1 | 6.8 | 8.2 | 5.5 | 3.3 | 2.6 | 7.0 |
| 2013 | 0.2 | 8.3 | 28.8 | 65.6 | 57.2 | 30.2 | 16.8 | 13.3 | 7.0 | 3.7 | 4.1 | 2.4 | 2.1 | 2.2 | 3.3 |
| 2014 | 1.5 | 21.5 | 28.5 | 38.1 | 64.1 | 53.7 | 21.7 | 12.1 | 8.7 | 4.0 | 2.9 | 2.6 | 1.6 | 2.1 | 2.9 |
| 2015 | 2.1 | 29.4 | 51.4 | 27.7 | 33.0 | 42.3 | 32.0 | 17.3 | 8.1 | 7.4 | 3.5 | 2.7 | 1.7 | 1.7 | 3.2 |
| 2016 | 5.2 | 20.4 | 59.2 | 67.4 | 37.2 | 30.3 | 29.5 | 23.1 | 11.3 | 9.4 | 5.3 | 2.7 | 2.3 | 1.3 | 5.1 |
| 2017 | 1.6 | 49.3 | 33.5 | 68.1 | 62.6 | 27.0 | 17.9 | 18.2 | 11.8 | 7.7 | 3.9 | 4.9 | 1.8 | 1.1 | 3.6 |
| 2018 | 0.0 | 0.8 | 18.5 | 49.9 | 32.0 | 29.2 | 20.9 | 11.4 | 8.9 | 11.0 | 7.5 | 6.2 | 3.2 | 2.2 | 3.7 |

Table 7.14.3. Weight-at-age for sole in 7.jk.

|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0.154 | 0.221 | 0.275 | 0.342 | 0.412 | 0.455 | 0.511 | 0.496 | 0.628 | 0.567 | 0.762 | 0.499 | 0.505 | 0.777 | 1.095 |
| 1994 | 0.143 | 0.233 | 0.278 | 0.346 | 0.421 | 0.453 | 0.514 | 0.552 | 0.610 | 0.632 | 0.632 | 0.583 | 0.660 | 0.845 | 0.661 |
| 1995 | 0.141 | 0.194 | 0.322 | 0.362 | 0.338 | 0.370 | 0.493 | 0.452 | 0.722 | 0.579 | 0.401 | 0.297 | 0.836 | 0.350 | 0.607 |
| 1996 | 0.138 | 0.169 | 0.230 | 0.307 | 0.435 | 0.421 | 0.505 | 0.587 | 0.613 | 0.712 | 0.755 | 0.643 | 0.765 | 0.723 | 0.673 |
| 1997 | 0.133 | 0.200 | 0.281 | 0.334 | 0.409 | 0.526 | 0.618 | 0.592 | 0.679 | 0.679 | 0.691 | 0.848 | 0.889 | 0.695 | 0.974 |
| 1998 | 0.136 | 0.223 | 0.281 | 0.357 | 0.379 | 0.448 | 0.515 | 0.554 | 0.455 | 0.647 | 0.497 | 0.641 | 0.659 | 0.763 | 0.819 |
| 1999 | 0.152 | 0.192 | 0.308 | 0.345 | 0.400 | 0.426 | 0.461 | 0.575 | 0.578 | 0.657 | 0.449 | 0.896 | 0.592 | 0.832 | 0.760 |
| 2000 | 0.180 | 0.210 | 0.255 | 0.396 | 0.416 | 0.472 | 0.503 | 0.489 | 0.506 | 0.452 | 0.555 | 0.818 | 0.525 | 0.850 | 0.694 |
| 2001 | 0.164 | 0.228 | 0.295 | 0.337 | 0.394 | 0.481 | 0.548 | 0.530 | 0.587 | 0.795 | 0.542 | 0.740 | 0.967 | 0.867 | 0.438 |
| 2002 | 0.203 | 0.198 | 0.254 | 0.305 | 0.469 | 0.490 | 0.473 | 0.654 | 0.730 | 0.721 | 0.626 | 0.616 | 1.150 | 0.643 | 0.871 |
| 2003 | 0.168 | 0.191 | 0.296 | 0.323 | 0.329 | 0.378 | 0.371 | 0.575 | 0.499 | 0.548 | 0.477 | 0.557 | 0.446 | 0.779 | 0.640 |
| 2004 | 0.094 | 0.199 | 0.197 | 0.293 | 0.313 | 0.353 | 0.287 | 0.584 | 0.636 | 0.499 | 0.595 | 0.499 | 0.845 | 0.457 | 0.761 |
| 2005 | 0.131 | 0.168 | 0.198 | 0.249 | 0.383 | 0.313 | 0.340 | 0.446 | 0.525 | 0.468 | 0.604 | 0.489 | 0.393 | 0.437 | 0.841 |
| 2006 | 0.160 | 0.180 | 0.205 | 0.257 | 0.298 | 0.354 | 0.354 | 0.377 | 0.456 | 0.377 | 0.612 | 0.438 | 0.568 | 0.508 | 0.775 |
| 2007 | 0.154 | 0.208 | 0.268 | 0.282 | 0.329 | 0.341 | 0.378 | 0.395 | 0.449 | 0.376 | 0.418 | 0.554 | 0.494 | 0.594 | 0.527 |
| 2008 | 0.144 | 0.204 | 0.236 | 0.278 | 0.305 | 0.339 | 0.339 | 0.395 | 0.389 | 0.445 | 0.560 | 0.450 | 0.512 | 0.457 | 0.744 |
| 2009 | 0.123 | 0.196 | 0.234 | 0.265 | 0.268 | 0.318 | 0.386 | 0.420 | 0.393 | 0.417 | 0.368 | 0.476 | 0.828 | 0.480 | 0.527 |


|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | 0.177 | 0.197 | 0.247 | 0.304 | 0.331 | 0.364 | 0.371 | 0.400 | 0.440 | 0.427 | 0.512 | 0.423 | 0.541 | 0.503 | 0.505 |
| 2011 | 0.186 | 0.207 | 0.236 | 0.260 | 0.298 | 0.340 | 0.420 | 0.479 | 0.469 | 0.523 | 0.580 | 0.600 | 0.597 | 0.485 | 0.639 |
| 2012 | 0.191 | 0.216 | 0.254 | 0.294 | 0.320 | 0.362 | 0.404 | 0.423 | 0.459 | 0.483 | 0.461 | 0.517 | 0.584 | 0.681 | 0.552 |
| 2013 | 0.141 | 0.226 | 0.268 | 0.302 | 0.339 | 0.352 | 0.404 | 0.440 | 0.483 | 0.483 | 0.546 | 0.614 | 0.477 | 0.557 | 0.647 |
| 2014 | 0.130 | 0.209 | 0.246 | 0.282 | 0.314 | 0.348 | 0.354 | 0.398 | 0.485 | 0.479 | 0.451 | 0.493 | 0.438 | 0.653 | 0.820 |
| 2015 | 0.152 | 0.206 | 0.231 | 0.284 | 0.316 | 0.319 | 0.330 | 0.374 | 0.393 | 0.455 | 0.476 | 0.533 | 0.404 | 0.643 | 0.510 |
| 2016 | 0.203 | 0.254 | 0.280 | 0.302 | 0.336 | 0.359 | 0.403 | 0.383 | 0.443 | 0.418 | 0.452 | 0.491 | 0.491 | 0.528 | 0.591 |
| 2017 | 0.144 | 0.204 | 0.236 | 0.278 | 0.305 | 0.339 | 0.339 | 0.395 | 0.389 | 0.445 | 0.560 | 0.450 | 0.512 | 0.457 | 0.693 |
| 2018 | 0.000 | 0.189 | 0.237 | 0.296 | 0.332 | 0.360 | 0.376 | 0.416 | 0.489 | 0.475 | 0.472 | 0.486 | 0.541 | 0.615 | 0.705 |

Table 39.4. Tuning data. The ages (3-9) and years used in the assessment are in bold.


## Table 7.14.5. XSA diagnostics.

```
FLR XSA Diagnostics 2019-05-26 22:06:56
CPUE data from indices
Catch data for 26 years 1993 to 2018. Ages 2 to 10.
    fleet first age last age first year last year alpha
beta
1 IRL-VMS: nos per 1000 hours 2 0 2006 2018 <NA>
<NA>
```

Time series weights :
Tapered time weighting not applied
Catchability analysis :
Catchability independent of size for all ages
Catchability independent of age for ages > 7
Terminal population estimation :
Survivor estimates shrunk towards the mean $F$
of the final 5 years or the 5 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1.5$
Minimum standard error for population
estimates derived from each fleet $=0.2$
prior weighting not applied
Regression weights
year
age 2009201020112012201320142015201620172018

| al1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Fishing mortalities
year
age $\begin{array}{lllllllllll}2009 & 2010 & 2011 & 2012 & 2013 & 2014 & 2015 & 2016 & 2017 & 2018\end{array}$
$2 \quad 0.000 \quad 0.008 \quad 0.002 \quad 0.002 \quad 0.000 \quad 0.003 \quad 0.004 \quad 0.015 \quad 0.086 \quad 0.000$
$3 \quad 0.037 \quad 0.075 \quad 0.044 \quad 0.034 \quad 0.025 \quad 0.040 \quad 0.056 \quad 0.050 \quad 0.172 \quad 0.051$

```
4 0.132 0.148 0.123 0.103 0.101 0.099 0.113 0.138 0.097 0.081
5 0.273 0.248 0.180 0.171 0.179 0.169 0.119 0.191 0.208 0.183
6 0.310 0.241 0.209 0.172 0.183 0.238 0.194 0.207 0.243 0.128
7 0.282 0.193 0.234 0.278 0.207 0.233 0.218 0.246 0.204 0.153
8 0.293 0.207 0.263 0.268 0.213 0.202 0.190 0.208 0.201 0.215
9 0.225 0.231 0.278 0.276 0.188 0.210 0.219 0.182 0.172 0.170
10 0.225 0.231 0.278 0.276 0.188 0.210 0.219 0.182 0.172 0.170
```

| XSA population number (Thousand) |  |  |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| age |  |  |  |  |  |  |  |  |  |
| year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 2009 | 778 | 434 | 326 | 337 | 124 | 72 | 69 | 83 | 119 |
| 2010 | 663 | 704 | 378 | 259 | 232 | 82 | 49 | 46 | 166 |
| 2011 | 398 | 595 | 591 | 295 | 183 | 165 | 61 | 36 | 168 |
| 2012 | 398 | 359 | 515 | 472 | 223 | 134 | 118 | 43 | 185 |
| 2013 | 642 | 359 | 314 | 420 | 360 | 170 | 92 | 82 | 152 |
| 2014 | 625 | 581 | 317 | 257 | 318 | 271 | 125 | 67 | 137 |
| 2015 | 492 | 564 | 505 | 260 | 196 | 227 | 195 | 92 | 151 |
| 2016 | 367 | 443 | 483 | 408 | 209 | 146 | 165 | 146 | 235 |
| 2017 | 21 | 327 | 381 | 380 | 305 | 154 | 104 | 121 | 231 |
| 2018 | 0 | 17 | 249 | 313 | 279 | 217 | 113 | 77 | 287 |

Estimated population abundance at 1st Jan 2019
age
$\begin{array}{llllllllll}\text { year } & 2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10\end{array}$
$20190015208236222168 \quad 8359$

Fleet: IRL-VMS: nos per 1000 hours

Log catchability residuals.

| age 20062007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 20172018 |  |  |  |  |  |  |  |  |  |
| 30.4830 .239 | 0.244 | -0.500 | 0.063 | -0.382 | -0.745 | -0.683 | -0.310 | 0.279 | 0.086 |
| 1.1810 .044 |  |  |  |  |  |  |  |  |  |
| $4 \quad 0.163 \quad 0.128$ | 0.619 | -0.019 | -0.060 | -0.153 | -0.432 | -0.072 | -0.199 | 0.173 | 0.305 |
| -0.197-0.259 |  |  |  |  |  |  |  |  |  |
| $5-0.3980 .056$ | 0.411 | 0.270 | 0.018 | -0.206 | $-0.361$ | 0.067 | -0.095 | -0.210 | 0.194 |
| 0.1350 .120 |  |  |  |  |  |  |  |  |  |
| $6-0.330 \quad 0.328$ | 0.172 | 0.288 | -0.116 | -0.167 | -0.464 | -0.023 | 0.136 | 0.172 | 0.167 |
| 0.179-0.343 |  |  |  |  |  |  |  |  |  |
| $7-0.3450 .088$ | 0.133 | 0.156 | -0.375 | -0.091 | -0.024 | 0.064 | 0.078 | 0.251 | 0.302 |
| -0.033-0.205 |  |  |  |  |  |  |  |  |  |

```
    8 -0.215 0.147 0.151 0.193 -0.307 0.027 -0.063 0.094 -0.070
-0.047 0.138
    9-0.009 0.169 0.071-0.070 -0.198 0.077 -0.028 -0.034 -0.032 0.256 0.007
-0.204 -0.098
```

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

```
Mean_Logq -0.0970 0.7042 1.1389 1.2470 1.2839 1.2839 1.2839
S.E_Logq 0.2749 0.2749 0.2749 0.2749 0.2749 0.2749 0.2749
```

Terminal year survivor and $F$ summaries:
, Age 3 Year class $=2015$
source
scaledwts survivors yrcls

| IRL-VMS: nos per 1000 hours | 0.877 | 15 | 2015 |
| :--- | :--- | :--- | :--- |
| fshk | 0.123 | 11 | 2015 |

, Age 4 Year class $=2014$
source

|  | scaledwts survivors | yrcis |  |
| :--- | ---: | ---: | ---: |
| IRL-VMS: nos per 1000 hours | 0.962 | 161 | 2014 |
| fshk | 0.038 | 152 | 2014 |

, Age 5 Year class $=2013$
source

|  | scaledwts survivors yrcls |  |  |
| :--- | ---: | ---: | ---: |
| IRL-VMS: nos per 1000 hours | 0.967 | 266 | 2013 |
| fshk | 0.033 | 250 | 2013 |

, Age 6 Year class $=2012$
source
scaledwts survivors yrcls

| IRL-VMS: nos per 1000 hours | 0.964 | 158 | 2012 |
| :--- | :--- | :--- | :--- |
| fshk | 0.036 | 128 | 2012 |

, Age 7 Year class $=2011$
source


Table 33.6. Summary table for sole 7.jk. Landings in tonnes (7.jk only). Recruitment (age 3) in thousands. SSB in tonnes.

| year | landings | recruit | fbar | ssb |
| :---: | :---: | :---: | :---: | :---: |
| 1993 | 246 | 920 | 0.357 | 699 |
| 1994 | 178 | 573 | 0.214 | 809 |
| 1995 | 241 | 943 | 0.383 | 686 |
| 1996 | 166 | 421 | 0.242 | 681 |
| 1997 | 191 | 663 | 0.298 | 681 |
| 1998 | 219 | 649 | 0.329 | 666 |
| 1999 | 200 | 575 | 0.707 | 549 |
| 2000 | 109 | 508 | 0.29 | 423 |
| 2001 | 127 | 749 | 0.225 | 462 |
| 2002 | 144 | 500 | 0.328 | 615 |
| 2003 | 102 | 561 | 0.377 | 433 |
| 2004 | 85 | 318 | 0.205 | 419 |
| 2005 | 77 | 247 | 0.184 | 358 |
| 2006 | 61 | 294 | 0.158 | 338 |
| 2007 | 83 | 562 | 0.216 | 371 |
| 2008 | 77 | 387 | 0.207 | 382 |
| 2009 | 69 | 434 | 0.188 | 375 |
| 2010 | 82 | 704 | 0.178 | 460 |
| 2011 | 87 | 595 | 0.139 | 511 |
| 2012 | 94 | 359 | 0.12 | 577 |
| 2013 | 91 | 359 | 0.122 | 584 |
| 2014 | 87 | 581 | 0.137 | 544 |
| 2015 | 78 | 564 | 0.121 | 541 |
| 2016 | 115 | 443 | 0.146 | 656 |
| 2017 | 114 | 327 | 0.18 | 550 |
| 2018 | 75 | 17 | 0.111 | 536 |
| 2019 |  | 553 |  | 460 |



Figure 33.1. The spatial distribution of International landings of sole (2012 data, all gears combined; data from STECF).


Figure 33.2. Irish OTB retained catches on observer trips in 7.jk during 2018. Numbers raised to fleet level using fishing effort (hours fished).

## SOL 7jk



Age

Figure 7.14.3. Age distribution of sole in 7.jk between 1993 and 2018. All gears and quarters combined.

## SOL 7jk

Standardised landings proportions-at-age


Figure 33.4. Standardised catch proportions-at-age for sole in 7.jk. Grey bubbles represent higher than average catch-at-age and black bubbles represent lower than average catch-at-age.


Figure 33.5. Catch weights/stock weights of sole 7.jk.


## 27.7j Sole



Figure 33.6. Top: the proportion of sole in landings of Irish vessels with VMS over the years 2006-2017. The black line indicates the polygon inside which sole are caught. Effort and landings from the VMS/logbooks data inside the polygon were used as a tuning index. Bottom: the VMS lpue index (black line) and the lpue of sole in the whole of 7.j (1995-2018).



Figure 33.7. The log-standardised tuning index by year (top) and cohort (bottom). The cohorts are tracked quite well and no year effects are obvious.

IRL-VMS: nos per 1000 hours


Figure 33.8. Internal consistency of the tuning fleet.

## Residuals

Sol 7jk
IRL-VMS: nos per 1000 hours


Figure 33.9. Residuals of the index fit.


Figure 33.10. Stock summary plot.

## Predictive distribution of recruitment for SOL7jk, WGCSE, COMBSEX, PLUSGROUP



Figure 33.11. Sole 7jk stock-recruit plot. Because recruitment does not appear to be impaired at the lowest stock size, the inflection point of the segmented regression was chosen to be Bloss. Estimated as part of WGCSE 2017.


Figure 33.12. Sole 7.jk Summary of MSY evaluations (without Btrigger harvest control rule), a) simulated and Marine 2018 observed recruitment, b) simulated and observed biomass, c) simulated and observed catch and d) Cumulative probability of $F_{m s y}$ and $S S B<B_{l i m}$ and $B_{p a}$.. Estimated as part of WGCSE 2017.


Figure 33.13. Sole 7.jk Summary of MSY evaluations (with $B_{\text {trigger }}=B_{\text {loss }}$ harvest control rule), a) simulated and observed recruitment, b) simulated and observed biomass, c) simulated and observed catch and d) Cumulative probability of $F_{\text {MSY }}$ and $S S B<B_{\lim }$ and $B_{\text {pa. }}$. Estimated as part of WGCSE 2017.

## 36 Whiting (Merlangius merlangus) in Division 6.a (West of Scotland)

## Type of assessment in 2019

An update/SPALY Time-Series Analysis (TSA) was carried out with catch and survey data, following the procedure outlined in the Stock Annex. No changes were considered with regard to reference points in relation to those estimated in the previous year.

## ICES advice applicable to 2019

ICES advises that when the precautionary approach is applied, there should be zero catch in each of the years 2019 and 2020.
http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2018/2018/whg.27.6a.pdf

## ICES advice applicable to 2017 and 2018

ICES advises that when the MSY approach is applied, there should be zero catch in each of the years 2017 and 2018.
http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2016/2016/whg-scow.pdf

### 36.1 General

## Stock description

General information is presented in the Stock Annex.

## Management applicable to 2018 and 2019

The TAC for whiting (in tonnes) is set for ICES subareas 6, 12 and 14 and EU and international waters of ICES Division 5b, for 2019 and 2018 is shown below.

TAC for 2019

| Species: | Whiting <br> Merlangius merlangus |  | Zone: | 6; Union and international waters of 5 b ; international waters of 12 and 14 <br> (WHG/56-14) |
| :---: | :---: | :---: | :---: | :---: |
| Germany |  | $3{ }^{1}{ }^{1}$ |  |  |
| France |  | $68\left({ }^{1}\right)$ |  |  |
| Ireland |  | $324{ }^{(1)}$ |  |  |
| United Kingdom |  | $717{ }^{(1)}$ |  |  |
| Union |  | $1112{ }^{(1)}$ |  |  |
| TAC |  | $1112{ }^{(1)}$ |  | Analytical TAC <br> Article 8 of this Regulation applies |

[^20]TAC for 2018

| Species:Whiting <br> Merlangius merlangus | Zone:6; Union and international waters of 5b; internat- <br> ional waters of 12 and 14 <br> (WHG/56-14) |  |
| :--- | ---: | :--- | :--- |
| Germany | $1\left(^{1}\right)$ |  |
| France | $26\left({ }^{1}\right)$ |  |
| Ireland | $64\left({ }^{1}\right)$ |  |
| United Kingdom | $122\left({ }^{1}\right)$ |  |
| Union | $213\left({ }^{1}\right)$ |  |
| TAC | $213\left({ }^{1}\right)$ |  |
|  |  |  |

(Council Regulation (EU) 2018/120).

The minimum landing size for whiting in Division 6.a is 27 cm .

## Fishery in 2018

A description of the fisheries in the West of Scotland is given in the Stock Annex.
Anecdotal information from the fishing industry suggests that the number of vessels targeting whiting continues to be very low. The year 2019 ends the transition period in the imposition of the landing obligation. The quota uptake for this stock in the preceding period varied from moderate to high. It is uncertain how the increased TAC will affect the catches of whiting.

Total landings (nominal landings, ICES statistics) in 2018 were 180 t , up by $6 \%$ from 2017 (Table 38.1). They were the second lowest in the time-series. The majority were landed by Scottish and Irish vessels, and smaller amounts by Danish, French and Dutch vessels. The UK landings in Division 6.a in 2018, constituted $83 \%$ of the pre-2019 quota for the UK, while Ireland exceeded its quota by $13 \%$. Total landings in 2018 constituted $84 \%$ of the TAC for that year.

The total estimated international catch of ages 1 and older in 2018 was 681 t of which 492 t were discards (Table 38.2). An additional 109 t were discarded as the 0 -group. Of the discards, $55 \%$ were discarded by the trawl fleet targeting crustaceans (Nephrops).
Mandatory introduction of larger square mesh panels for the Nephrops fleet in 2008 does not seem to have had much of an effect on the discards of whiting in Division 6.a in 2018. In terms of quantity, the discards in 2018 (ages 1 and older) were by $41 \%$ lower than those in 2017 and also below the average in the last decade. In terms of discard rate (discards as a proportion of catch), they were the tenth highest in the time-series.

The general perception from fishermen is that large number of whiting are being discarded by the Nephrops fleet and that the numbers of smaller whiting has exploded recently but mainly in inshore areas.

### 36.2 Data

## Landings

Total landings, as officially reported to ICES in 1965-2018, are shown in Figure 38.1 and Table 38.2. In the past, there had been concerns that the quality of landings data was deteriorating, giving a possible reason for the different stock dynamics implied by the commercial fleet and the annual survey (ScoGFS-WIBTS-Q1) being in operation at that time (see Section 5.1.6.1.3 in the 2005 WG Report; ICES, 2005). As a result, the total landings data from 1995 to 2005 are not used
in the assessment. Improved compliance measures and the introduction of UK and Irish legislation requiring registration of all fish buyers and sellers ensures that the reported total landings from 2006 onwards are more representative of actual landings.

Landings uploaded to InterCatch by métier and country are shown in Figure 38.2. Age distributions were estimated from market samples. Annual numbers-at-age in the landings are given in Table 38.3. Annual mean weights-at-age in the landings are given in Table 38.6 and shown in Figure 38.3. They decreased considerably last year in nearly all age groups. Overall, the mean weights-at-age in the landings have been variable in recent years due to the variability associated with low sample sizes. Efforts to increase sampling in these fisheries are being pursued.

## Discards

This year, WG estimates of discards are based on data collected in the Irish and Scottish discard programme (raised by weighted average to the level of the total international discards). Discard age compositions from Scottish and Irish samples have been applied to unsampled fleets. Discards uploaded to InterCatch by métier and country are shown in Figure 38.2.

Annual numbers-at-age in the discards are given in Table 38.4. Annual mean weights-at-age in the discards are given in Table 38.7 and shown in Figure 38.3.

## Biological

Annual numbers-at-age in the total catch are given in Table 38.5. Annual mean weights-at-age in the total catch are given in Table 38.8 and shown in Figure 38.3. As in previous assessments, the catch mean weights-at-age were also used as stock mean weights-at-age (see the Stock Annex).
Natural mortality $(M)$ is assumed to vary and be dependent on fish weight (Lorenzen, 1996). $M$ values are time-invariant and are calculated as:

$$
M_{a}=3.0 \bar{W}_{a}^{-0.29}
$$

where $M_{a}$ is natural mortality-at-age $a, \bar{W}_{a}$ is the time averaged stock weight-at-age $a$ (in g ) and the numbers are the Lorenzen's parameters for fish in natural ecosystems.
Maturity-at-age was assumed to be knife-edge, with the value 0 at age 1 and full maturity-at-age 2+ according to the Stock Annex.

## Surveys

Five research vessel survey series for whiting in 6 .a were available to the WG.

- Scottish first-quarter west coast groundfish survey (ScoGFS-WIBTS-Q1): all ages 1 and older, years 1985-2010;
- Scottish fourth-quarter west coast groundfish survey (ScoGFS-WIBTS-Q4): all ages including age 0, years 1996-2009.

The Q1 Scottish Groundfish survey was running in the period 1981-2010, and this was performed using a repeat station format with the GOV survey trawl together with the west coast groundgear rig, 'C'. Similarly, the Q4 Scottish Groundfish survey was running in 1996-2009, once again using the GOV survey trawl with groundgear ' $C$ ' and the fixed station format. The Q4 survey was not carried out in 2010 due to an engine break down of the research vessel.

In 2011, the Q1 and Q4 Scottish Groundfish surveys were re-designed. The previous repeat station survey format consisting of the same series of survey trawl positions being sampled at approximately the same temporal period every year is considered a rather imprecise method for
surveying both these subareas. Therefore, a move towards some sort of random stratified survey design was judged necessary (see further details of the modified survey design in the Stock Annex). The introduction of the new design initiated two time-series:

- Scottish first-quarter west coast groundfish survey (UK-SCOWCGFS-Q1): all ages 1 and older, years 2011-2019;
- Scottish fourth-quarter west coast groundfish survey (UK-SCOWCGFS-Q4): all ages including age 0, years 2011-2018.
The distribution of whiting at-age (cpue) in the Q1 and Q4 surveys in 2016-2019 is shown in Figure 38.4. The Q4 survey in 2013 (not shown in Figure 38.4) was not fully implemented due to adverse weather conditions; it covered only the northern half of Division 6.a and is therefore not used in the assessment. The Q1 survey in 2019 has recently been completed and processed. As a result, nine years of data are currently available in the time-series for the Q1 survey and seven years of data for the Q4 survey (as valid indices). These data were made available in this year's assessment.

The Irish groundfish survey:

- Irish fourth-quarter west coast groundfish survey (IGFS-WIBTS-Q4): all ages including age 0 , years 2003-2018.

The distribution of whiting at-age (cpue) in the two Q4 surveys, UKSGFS-WIBTS-Q4 (only the southern part) and IGFS-WIBTS-Q4 in 2015-2018 is shown in Figure 38.5. The previous Irish survey (IreGFS), being in operation in 1993-2002 (see the Stock Annex), is not used in the assessment. The current Irish survey uses the RV Celtic Explorer and is part of the IBTS coordinated western waters surveys. The vessel uses a GOV trawl, and the design is a depth stratified survey with randomised stations. Effort is recorded in terms of minutes towed. This time-series was considered long enough to be used in the assessment of whiting in Division 6.a, giving useful additional indications of year-class strength.

Further descriptions of the above five surveys can be found in "Manual of the IBTS North Eastern Atlantic Surveys" (ICES, 2017) and in the last IBTSWG report (ICES, 2018).

During the Inter-Benchmark Protocol of West of Scotland Roundfish (IBPWSRound), it was decided to include the new Scottish survey time-series in the assessment (ICES, 2015). An attempt was made to use one index to represent the stock abundance combining the two Q4 surveys currently in operation, IGFS-WIBTS-Q4 and UK-SCOWCGFS-Q4. However, considerable differences were found between the two surveys with cpue being overall higher in the Irish survey. As a consequence of these differences, the IBPWSRound agreed to continue using the Irish Q4 survey as an independent time-series. The differences between the two Q4 surveys were explored in more detail by Working Group on Improving use of Survey Data for Assessment and Advice (WGISDAA) in 2018. The group's interim report will soon be available. However, the issue will further explored at WGISDAA in 2019. Ultimately, five survey time-series were used in the present assessment of the whiting stock (just as last year).

The five surveys indices are shown in Table 38.9 with the data used in the final assessment being highlighted in bold.

A comparison of scaled (standardised to z-scores) survey indices (from the five time-series) atage show roughly similar trends, mainly for the Scottish surveys, for most ages (up to age 5, Figure 38.6). The two new Scottish surveys seem to show greater consistency (on a year basis) compared to the previous surveys.

Log mean-standardised survey indices by year class and by year in the Irish survey and new Scottish time-series are shown in Figure 38.7. Given the short length of the survey time-series,
the year-class plots demonstrate, in most cases, the ability of the surveys to reliably track year classes and to identify the stronger/weaker than average year classes.

The $\log$ catch curves for the commercial catch and for the surveys are shown in Figure 38.8. The curves for both ScoGFS-WIBTS-Q1 and ScoGFS-WIBTS-Q4 (the two surveys discontinued in 2010 in 2009, respectively) are relatively linear and not very noisy. They also show a fairly steep and consistent drop in abundance. Patterns are less clear with the Irish survey. Little can be said in this respect about the new survey time-series (UK-SCOWCGFS-Q1 and UK-SCOWCGFS-Q4) as they are relatively short.

## Commercial cpue

Four commercial catch-effort time-series were previously available to the WG, but they have not been used for a number of years. They are only presented in the Stock Annex.

### 36.3 Historical stock development

The final assessment of whiting in 6.a was conducted using a TSA model. The method was first developed by Gudmundsson (1994), and it was modified by Rob Fryer for the purpose of assessing time-series containing several years with survey data but no reliable catch data (Fryer, 2002). Subsequent enhancements to the method are detailed in Needle and Fryer (2002). The TSA model allows for years with missing catch or survey data.
Alternative exploratory assessments conducted using SURBA (Needle, 2003) and a Bayesian approach (Cook, 2012) were presented at the WKROUND benchmark in 2012 (ICES, 2012), but were not further explored in this assessment.

## Data screening and exploratory runs

Model used: TSA
Software used: NAG library (FORTRAN DLL) and functions in R.
Input data types and characteristics:

- Landings, ages 1-7+, years 1981-2018 (1995-2005 age structure only used),
- Discards, ages 1-7+, years 1981-2018 (1995-2005 age structure only used)
- ScoGFS-WIBTS-Q1, ages 1-6, years 1985-2010
- ScoGFS-WIBTS-Q4, ages 1-6, years 1996-2009
- IGFS-WIBTS-Q4, ages 1-4, years 2003-2006 and 2008-2018
- UK-SCOWCGFS-Q1: ages 1-6, years 2011-2019
- UK-SCOWCGFS-Q4: ages 1-6, years 2011-2012 and 2014-2018

The assessment of whiting in 6.a was conducted using a TSA model with updated survey data (five time-series). The details of the method are presented in the Stock Annex. No modification to the landings was made to account for area misreporting although total landings are excluded from the assessment for the years 1995-2005 as the reported landings data are considered to be unreliable during this period. (ICES, 2012). A "hockey-stick" model was employed to describe the stock-recruitment relationship. Some extra variability in landings and discards was allowed for some ages. Also, some points in the time-series that were identified as outliers were downweighted to improve the fit. One point in the IGFS-WIBTS-Q4 time-series (for 2007) was treated as an outlier and was excluded from the analysis. Similarly, one point in UK-SCOWCGFS-Q4 (for 2013) was excluded as the survey was not complete in that year. Table 38.10 shows the TSA parameter settings for the assessment run.

The main diagnostics of the quality of the model fit was the value of the objective function $\left(-2^{*} \log\right.$ likelihood), residuals and a consideration of how well the model has replicated discard ratios in the input data.

The WG assessment in 2015 was not properly optimised. The introduction of the new survey time-series at IBPWSRound had a considerable effect (not anticipated at that time) on some of the model parameters. In the 2016 assessment, greater care was taken to ensure that the model parameters were accurately chosen, which consequently improved the model's performance. That alteration resulted in a downward revision of the stock biomass compared to the 2015 assessment. This year's assessment closely followed the optimisation setup used in the last three assessments.

At IBPWSRound, TSA runs with and without a survey catchability trend were compared (ICES, 2015). In the latter, the parameters for persistent and transitory trends in survey catchability were both set to 0 . Given the overestimation of catch and uncertainty in the assessment with fixed survey catchability, this option was not further explored and the assessment including estimation of survey catchability trend was retained, which also applied from 2016 onwards.

## Final assessment

The TSA run using the five surveys is presented as the final assessment run. Table 38.11 shows the TSA parameter estimates for the assessment.

Figure 38.9 shows the proportion discarded at-age from the final TSA run. Discards continue to account for a large proportion of the total catch, with no obvious tendency to decrease or to level off.

Table 38.12 gives the TSA population numbers-at-age and Table 38.13 gives the associated standard errors. Estimated F at-age is given in Table 38.14 and standard errors on the log of this mortality are given in Table 38.15. Full summary output is given in Table 38.16.

Standardised residuals for landings and discards are given in Figure 38.10, and those for the five surveys in Figure 38.11. None of these are large enough to invalidate the model fit and there are no obvious time-trends in recent years.

TSA also estimated a change in catchability (this is plotted as the percentage change compared to the catchability at the start of each of the five surveys, Figure 38.12). There was a large increase in catchability in the two previous Scottish surveys and in the Irish current survey (although in recent years, this trend seems to have been reversed for the Irish survey). No major change could be seen in the new Scottish surveys in this respect. In fact, the parameters for persistent trends in survey catchability in the Q4 survey was found to be 0 in this assessment (Table 38.11).

The TSA stock-recruit plot is presented in Figure 38.13 and shows a rather good relationship, partly because the stock was driven to very low levels of SSB in 2006-2010. Both SSB and recruitment have been at their highest levels in the two last decades. The summary plots for the final assessment are shown in Figure 38.14.

The final estimates for the stock are:

$$
\begin{aligned}
& \mathrm{F}_{(2-4)} \text { in } 2018=0.032 \\
& \text { SSB in } 2019=18961 \mathrm{t}
\end{aligned}
$$

Retrospectives for the final assessment run are shown in Figure 38.15. This figure also shows lines at $\pm 2$ se (approximate $95 \%$ confidence limits) around the run in the last year. Retrospective bias is small with respect to SSB. With respect to mean F and recruitment, the results are roughly
within the confidence limits of this year's run. The confidence interval for mean F reflects uncertainty in estimation of mean F when that estimation is based to a large extent on survey data (1995-2005) or the age structure of discards data (2006 onwards).

## Comparison with last year's assessment

The above estimates show relatively high consistency with regard to F with the last year's assessment. The SSB estimate for 2018 was revised down by 6 thousand tonnes in this assessment.
$\mathrm{F}_{(2-4)}$ in $2017=0.041$ (the present assessment: in 2017, 0.043)
SSB in 2018= 23143 t (the present assessment: in 2018, 17175 t )

## State of the stock

The spawning-stock biomass (SSB) has been increasing since 2010 but remains very low compared to the historical estimates and is below Blim. Fishing mortality ( F ) has declined continuously since around 2000 and is now well below $\mathrm{F}_{\text {mSY. Recruitment is estimated to have been very }}$ low since 2002 but estimated to have increased in recent years.

### 36.4 Short-term projections

No short-term projection was conducted this year as the forecast for this stock is up-dated biennially starting from the year 2018. The next forecast is scheduled in 2020.

The last short-term projection (in 2018) followed the procedure outlined in the Stock Annex.

### 36.5 MSY explorations

The reference points for this stock were not updated in this assessment.
In 2016, MSY reference points and ranges were calculated for the stock using the same procedure as that agreed at WKMSYREF4 (ICES, 2016a). The details of the analysis and the results are presented in Working Document 7 (ICES, 2016b).

### 36.6 MSY and Biological reference points

The reference points estimated in 2016 are summarised in the table below:

| Reference point | WKMSY-REF4 <br> 2016 | WGCSE 2016 | Rationale (WKSYREF4) |
| :--- | :---: | :---: | :--- |
| $\mathrm{B}_{\text {lim }}$ | 28500 t | 31900 t | SSB value at the change point in the segmented regression |
| $\mathrm{B}_{\mathrm{pa}}$ | 39900 t | 44600 t | $\mathrm{B}_{\text {lim }} \times 1.4$ |
| $\mathrm{~F}_{\text {lim }}$ | 0.25 | 0.27 | Based on segmented regression simulation of recruitment |
| $\mathrm{F}_{\mathrm{pa}}$ | 0.18 | 0.19 | $\mathrm{~F}_{\text {lim }} / 1.4$ |
| $\mathrm{~F}_{\text {MSY }}$ | 0.22 | 0.23 | with $\mathrm{B}_{\text {trigger }}\left(=\mathrm{B}_{\mathrm{pa}}\right)$ |
|  | 0.16 | 0.18 | upper precautionary with $\mathrm{B}_{\text {trigger }}\left(=\mathrm{B}_{\mathrm{pa}}\right)$ |
| $\mathrm{F}_{\text {MSY }}$ upper | 0.34 | 0.32 | with $\mathrm{B}_{\text {trigger }}\left(=\mathrm{B}_{\mathrm{pa}}\right)$ |
| $\mathrm{F}_{\text {MSY }}$ lower | 0.16 | 0.15 | with $\mathrm{B}_{\text {trigger }}\left(=\mathrm{B}_{\mathrm{pa}}\right)$ |
| MSY $\mathrm{B}_{\text {trigger }}$ | 39900 t | 44600 t | $\mathrm{B}_{\mathrm{pa}}$ |
| Median SSB at | 36600 |  |  |

### 36.7 Management plans

There are no specific management objectives or a management plan for this stock, but a plan is under development.

### 36.8 Uncertainties and bias in the assessment and forecast

The most significant problem with assessment of this stock is related to the commercial data. Incorrect reporting of landings (species and quantity) is known to have occurred in the past and directly affecting the perception of the stock. TSA is explicitly designed to allow omission of catch data (1995-2005 uses only age structure data from the catch), which is why it was used here to run the final assessment.

The survey data and commercial catch data contain different signals concerning the stock. A similar problem has been present in the North Sea whiting stock (as reported by ICES, 2010). Three potential sources of this discrepancy were identified for the North Sea stock, and they may apply to whiting in 6.a as well: bias in catch estimates, changes in survey catchability or changes in natural mortality due to predation or regime shift (ICES, 2010). Allowing the TSA assessment to interpret this difference as a persistent trend (increase) in survey catchability may lead to an underestimation of stock size, but the magnitude of underestimation is unknown.

After being explored extensively, new reliable reference points were eventually delivered by the 2016 WG for the stock and, if necessary, will be updated in future assessments.

Long-term information on the historical yield and catch composition indicates that the present stock size is low. The current assessment also indicates that the stock is at a low level. Total mortality has been declining over the past few years. The sum of the Scottish West Coast groundfish survey indices, (both in quarter one and quarter four) is also low, but shows a moderate increase from 2010 onwards.

### 36.9 Recommendation for next benchmark

A landings and discards disaggregated assessment appeared to be a reliable basis for determining the status of the whiting stock in Division 6.a.

The emergence of a trend in survey catchability needs to be addressed. The cause of this is very uncertain. Trends in catchability have been a feature of this assessment in the past and point to some issues with the model structure or assumptions. There have been significant changes in the commercial fishing practices in recent years that are not explicitly taken into account by this assessment model (e.g. emergency measures since 2010 and decline in the TR1 gadoid fishery prior to that). This will require detail explorations in the next benchmark.

The discrepancy in catch rate between the two Q4 surveys, IGFS-WIBTS-Q4 and UKSGFS-WI-BTS-Q4 is subject to further exploration (by WGISDAA). This analysis may provide, especially with more years of data available, more insight into the difference in catchability in the two surveys with the ultimate goal of creating one common index.

With regard to the assessment method, changes to the variance structures used in the model should be allowed if they improve model diagnostics (e.g. likelihood ratio tests, prediction error and residual plots).

### 36.10 Management considerations

Recruitment during the 1990s appears to have been high while after the year 2000, it has been below average. A number of relatively strong (compared to the recent past) year classes have been recorded recently (fish recruited in 2015-2019).

Whiting are caught in mixed fisheries with cod and haddock in Division 6.a. Management of whiting will be strongly linked to that for cod for which there is an ongoing recovery plan (EC, 2008). There have also been several technical conservation measures introduced in the 6 .a gadoid fishery in recent years including the mandatory increases in mesh size to 120 mm .

Whiting are caught and heavily discarded in small-meshed fisheries for Nephrops. When this stock falls under the landing obligation, it can (in the presence of high discards and low quota) become a "choke species" for the Nephrops fishery.

### 36.11 References

Cook, R. M. 2012. Assessment of West of Scotland cod (ICES Division VIa) using a Bayesian approach, Working paper to WKROUND 2012, February 2012: 1-31.

EC. 2008. Council Regulation (EC) No 1342/2008 of 18 December 2008 establishing a long-term plan for cod stocks and the fisheries exploiting those stocks and repealing, Regulation (EC) No 423/2004, OJ L 348, 24.12.2008: 20-33.

Fryer, R. J. 2002. TSA: is it the way? Appendix D in report of Working Group on Methods on Fish Stock Assessment. ICES CM 2002/D:01.

Gudmundsson, G. 1994. Time-series analysis of catch-at-age observations. Applied Statistics 43: 117-126.
ICES. 2005. Report of the Working Group on the Assessment of Northern Shelf Demersal Stocks (WGNSDS), 10-19 May 2005, Murmansk, Russia, ICES CM 2005/ACFM:13, pp. 644.
ICES. 2010. Report of the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), 5-11 May 2010, ICES Headquarters, Copenhagen, ICES CM 2010/ACOM:13, pp. 1072.

ICES. 2012. Report of the Benchmark Workshop on Western Waters Roundfish (WKROUND), 22-29 February 2012, Aberdeen, UK. ICES CM 2012/ACOM:49, pp. 283.
ICES. 2015. Report of the Inter-Benchmark Protocol of West of Scotland Roundfish (IBPWSRound), Febru-ary-April 2015, By correspondence. ICES CM 2015/ACOM:37. 72 pp.

ICES. 2016a. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 183 pp.

ICES. 2016b. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 4-13 May 2016, Copenhagen, Denmark. ICES CM 2016/ACOM:13. 1343 pp.

ICES. 2017. Manual of the IBTS North Eastern Atlantic Surveys. Series of ICES Survey Protocols SISP 15. 92 pp. http://doi.org/10.17895/ices.pub. 3519.

ICES. 2018. Report of the International Bottom Trawl Survey Working Group (IBTSWG), 19-23 March 2018, Oranmore, Ireland. ICES CM 2018/EOSG:01. 233 pp.
Lorenzen, K. 1996. The relationship between body weight and natural mortality in juvenile and adult fish: a comparison of natural ecosystems and aquaculture. Journal of Fish Biology, 49: 627-647.

Needle, C. L. 2003. Survey-based assessments with SURBA. Working Document to the ICES Working Group on Methods of Fish Stock Assessment, Copenhagen, 29 January-5 February 2003.

Needle, C. L. and Fryer, R. J. 2002. A modified TSA for cod in Division VIa: separate landings and discards. Working document to the ICES Advisory Committee on Fisheries Management, October 2002.

Table 38.1. Whiting in Division 6.a. Nominal landings (in tonnes) as officially reported to ICES.

| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 1 | - | + | - | + | + | + | - | 1 | 1 | + | - | - | - | - | + | - | - | - | - |
| Denmark | 1 | + | 3 | 1 | 1 | + | + | + | + | - | - | - | - | - | + | + | - | - | - | - |
| Faroe Islands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | + | + |
| France | 199 | 180 | 352 | 105 | 149 | 191 | 362 | 202 | 108 | 82 | 300 | 48 | 52 | 21 | 11 | 6 | 9 | 7 | 6 | 1 |
| Germany | + | + | + | 1 | 1 | + | - | + | - | - | + | - | - | - | - | - | - | + | 1 | - |
| Ireland | 1315 | 977 | 1200 | 1377 | 1192 | 1213 | 1448 | 1182 | 977 | 952 | 1121 | 793 | 764 | 577 | 568 | 356 | 172 | 196 | 56 | 69 |
| Netherlands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Norway | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Spain | - | - | - | - | - | - | 1 | - | 1 | 2 | + | - | 2 | - | - | - | - | - | - | - |
| UK (E, W and NI) | 44 | 50 | 218 | 196 | 184 | 233 | 204 | 237 | 453 | 251 | 210 | 104 | 71 | 73 | 35 | 13 | 5 | 2 | 1 |  |
| UK (Scot.) | 6109 | 4819 | 5135 | 4330 | 5224 | 4149 | 4263 | 5021 | 4638 | 3369 | 3046 | 2258 | 1654 | 1064 | 751 | 444 | 103 | 178 | 424 |  |
| UK (total) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 370 |
| Total landings | 7669 | 6026 | 6908 | 6010 | 6751 | 5786 | 6278 | 6642 | 6178 | 4657 | 4677 | 3203 | 2543 | 1735 | 1365 | 819 | 289 | 383 | 488 | 441 |

Table 38.1. Continued.

| Country | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017* | 2018* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | - | - | - | - | - | - | - | - | - | - |
| Denmark | - | - | - | - | - | - | - | - | - | 2 |
| Faroe Islands | - | + | 1 | 1 | - | - | - | - | - | - |
| France | 1 | 3 | + | + | 1 | 1 | + | 5 | 3 | 2 |
| Germany | - | - | - | - | - | - | - | - | - | - |
| Ireland | 125 | 99 | 149 | 96 | 97 | 97 | 88 | 77 | 53 | 72 |
| Netherlands | - | - | - | - | - | - | 11 | 52 | 19 | 2 |
| Norway | 2 | - | - | - | - | - | - | - | - | + |
| Spain | - | - | - | - | - | - | - | - | - | - |
| UK (E, W and NI) |  |  |  |  |  |  |  |  |  |  |
| UK (Scot.) |  |  |  |  |  |  |  |  |  |  |
| UK (total) | 354 | 247 | 80 | 204 | 116 | 83 | 122 | 98 | 94 | 101 |
| Total landings | 482 | 349 | 230 | 301 | 214 | 181 | 221 | 232 | 169 | 180 |

* Preliminary.
$+<0.5$ t.

Table 38.2. Whiting in Division 6.a. Landings, discards and catch estimates 1978-2018, as used by the WG. Values are totals for fish over the ages 1 to 7+. Discard and catch values for the years 1978-2003 are revised compared to previous assessments because of a revised method for raising discards.

| Year | Weight (tonnes) |  | Discards | Numbers (thousands) |  | Discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total | Human consumption |  | Total | Human consumption |  |
| 1978 | 19346 | 14677 | 4669 | 85502 | 54369 | 31133 |
| 1979 | 20100 | 17081 | 3019 | 77484 | 61393 | 16091 |
| 1980 | 14598 | 12816 | 1782 | 54643 | 44562 | 10081 |
| 1981 | 14335 | 12203 | 2132 | 59247 | 46067 | 13180 |
| 1982 | 19356 | 13871 | 5485 | 84886 | 47883 | 37003 |
| 1983 | 22264 | 15970 | 6294 | 86244 | 49359 | 36885 |
| 1984 | 20475 | 16458 | 4017 | 89113 | 50218 | 38895 |
| 1985 | 17733 | 12893 | 4840 | 75192 | 43166 | 32026 |
| 1986 | 11123 | 8454 | 2669 | 49413 | 31273 | 18140 |
| 1987 | 23462 | 11544 | 11918 | 158176 | 41221 | 116955 |
| 1988 | 19484 | 11352 | 8132 | 109474 | 40681 | 68793 |
| 1989 | 13407 | 7531 | 5876 | 72364 | 26876 | 45488 |
| 1990 | 10173 | 5643 | 4530 | 51426 | 19201 | 32225 |
| 1991 | 11543 | 6660 | 4883 | 63767 | 25103 | 38664 |
| 1992 | 15253 | 6004 | 9249 | 93424 | 22266 | 71158 |
| 1993 | 11631 | 6872 | 4759 | 52365 | 23246 | 29119 |
| 1994 | 9356 | 5901 | 3455 | 44986 | 20060 | 24926 |
| 1995 | 11847 | 6076 | 5771 | 66432 | 18763 | 47669 |
| 1996 | 15096 | 7156 | 7940 | 81230 | 22329 | 58901 |
| 1997 | 11536 | 6285 | 5251 | 55724 | 19250 | 36474 |
| 1998 | 13847 | 4631 | 9216 | 88803 | 14387 | 74416 |
| 1999 | 8588 | 4613 | 3975 | 43219 | 15970 | 27249 |
| 2000 | 16295 | 3010 | 13285 | 176734 | 10118 | 166616 |
| 2001 | 6701 | 2438 | 4263 | 38114 | 8477 | 29637 |
| 2002 | 4560 | 1709 | 2851 | 28381 | 5765 | 22616 |
| 2003 | 2075 | 1356 | 719 | 10063 | 4124 | 5939 |


| Year | Weight (tonnes) |  | Discards | Numbers (thousands) |  | Discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total | Human consumption |  | Total | Human consumption |  |
| 2004 | 3437 | 811 | 2626 | 21749 | 2571 | 19178 |
| 2005 | 1239 | 341 | 898 | 6154 | 1051 | 5103 |
| 2006 | 1326 | 380 | 946 | 12988 | 1049 | 11939 |
| 2007 | 849 | 484 | 365 | 4879 | 1145 | 3734 |
| 2008 | 617 | 443 | 174 | 3085 | 1232 | 1853 |
| 2009 | 905 | 488 | 417 | 18038 | 1115 | 16923 |
| 2010 | 1193 | 307 | 886 | 18391 | 601 | 17790 |
| 2011 | 569 | 230 | 339 | 4877 | 583 | 4294 |
| 2012 | 1041 | 313 | 729 | 9679 | 702 | 8977 |
| 2013 | 1175 | 222 | 953 | 15444 | 522 | 14922 |
| 2014 | 770 | 184 | 586 | 11226 | 408 | 10818 |
| 2015 | 1060 | 227 | 833 | 9336 | 479 | 8857 |
| 2016 | 1029 | 233 | 796 | 7102 | 433 | 6669 |
| 2017 | 1386 | 176 | 1209 | 11572 | 386 | 11187 |
| 2018 | 681 | 188 | 492 | 4491 | 433 | 4058 |
| Min | 569 | 176 | 174 | 3085 | 386 | 1853 |
| GM | 5062 | 2243 | 2287 | 314598 | 6704 | 20101 |
| AM | 9304 | 5567 | 3737 | 50132 | 18996 | 31136 |
| Max | 23462 | 17081 | 13285 | 176734 | 61393 | 166616 |

Table 38.3. Whiting in Division 6.a. Landings-at-age (thousands).

|  | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1965 | 6938 | 6085 | 43530 | 4803 | 388 | 103 | 22 |
| 1966 | 1685 | 10544 | 2229 | 28185 | 1861 | 186 | 52 |
| 1967 | 5169 | 26023 | 10619 | 697 | 14574 | 789 | 143 |
| 1968 | 7265 | 16484 | 9239 | 3656 | 324 | 5036 | 368 |
| 1969 | 873 | 25174 | 8644 | 2566 | 1206 | 118 | 2333 |
| 1970 | 730 | 6423 | 28065 | 3241 | 670 | 214 | 550 |
| 1971 | 2387 | 8617 | 4122 | 34784 | 1338 | 240 | 223 |
| 1972 | 16777 | 12028 | 4013 | 1363 | 14796 | 793 | 148 |
| 1973 | 14078 | 36142 | 5592 | 1461 | 357 | 4292 | 310 |
| 1974 | 9083 | 51036 | 10049 | 1166 | 180 | 52 | 849 |
| 1975 | 14917 | 16778 | 36318 | 2819 | 281 | 57 | 245 |
| 1976 | 8500 | 46421 | 15757 | 17423 | 1508 | 66 | 57 |
| 1977 | 16120 | 13376 | 25144 | 3127 | 4719 | 292 | 24 |
| 1978 | 17670 | 18175 | 6682 | 9400 | 941 | 1433 | 68 |
| 1979 | 6334 | 34221 | 13282 | 3407 | 3488 | 276 | 384 |
| 1980 | 11650 | 11378 | 14860 | 4155 | 1244 | 1085 | 190 |
| 1981 | 3593 | 24395 | 11297 | 4611 | 1518 | 452 | 201 |
| 1982 | 2991 | 5783 | 29094 | 6821 | 2043 | 803 | 348 |
| 1983 | 3418 | 7094 | 8040 | 22757 | 6070 | 1439 | 540 |
| 1984 | 7209 | 12765 | 8221 | 4387 | 14825 | 1953 | 858 |
| 1985 | 4139 | 19520 | 8574 | 3351 | 1997 | 4764 | 822 |
| 1986 | 2674 | 14824 | 9770 | 2653 | 532 | 291 | 529 |
| 1987 | 6430 | 13935 | 13988 | 5442 | 837 | 330 | 259 |
| 1988 | 1842 | 20587 | 9638 | 6168 | 1949 | 290 | 207 |
| 1989 | 2529 | 5887 | 11889 | 4767 | 1266 | 468 | 71 |
| 1990 | 3203 | 8028 | 2393 | 4009 | 1326 | 204 | 37 |
| 1991 | 3294 | 8826 | 10046 | 1208 | 1391 | 286 | 51 |


| Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1992 | 2695 | 9440 | 4473 | 4782 | 396 | 373 | 106 |
| 1993 | 1051 | 10179 | 6293 | 2673 | 2738 | 163 | 147 |
| 1994 | 909 | 4889 | 9158 | 3607 | 712 | 715 | 69 |
| 1995 | 215 | 4322 | 6516 | 5654 | 1397 | 376 | 282 |
| 1996 | 990 | 5410 | 7675 | 5052 | 2461 | 583 | 157 |
| 1997 | 877 | 3658 | 8514 | 4316 | 1441 | 338 | 106 |
| 1998 | 840 | 3504 | 4277 | 3698 | 1442 | 338 | 288 |
| 1999 | 1013 | 6131 | 4546 | 2040 | 1774 | 355 | 112 |
| 2000 | 484 | 2952 | 4211 | 1570 | 485 | 328 | 89 |
| 2001 | 461 | 3271 | 2630 | 1567 | 401 | 131 | 16 |
| 2002 | 62 | 1624 | 3018 | 799 | 227 | 23 | 13 |
| 2003 | 170 | 710 | 1111 | 1673 | 347 | 111 | 2 |
| 2004 | 54 | 724 | 543 | 521 | 622 | 78 | 29 |
| 2005 | 28 | 276 | 455 | 140 | 99 | 45 | 7 |
| 2006 | 82 | 139 | 369 | 260 | 61 | 113 | 24 |
| 2007 | 187 | 168 | 255 | 326 | 132 | 27 | 50 |
| 2008 | 6 | 265 | 394 | 336 | 152 | 55 | 24 |
| 2009 | 59 | 216 | 254 | 430 | 100 | 44 | 13 |
| 2010 | 53 | 94 | 153 | 119 | 126 | 24 | 31 |
| 2011 | 0 | 310 | 133 | 82 | 28 | 17 | 12 |
| 2012 | 9 | 25 | 375 | 210 | 57 | 15 | 11 |
| 2013 | 21 | 49 | 83 | 277 | 67 | 18 | 7 |
| 2014 | 12 | 30 | 131 | 102 | 99 | 23 | 11 |
| 2015 | 11 | 83 | 61 | 164 | 69 | 67 | 25 |
| 2016 | 1 | 73 | 166 | 75 | 74 | 16 | 28 |
| 2017 | 16 | 35 | 167 | 71 | 49 | 38 | 10 |
| 2018 | 0 | 71 | 88 | 199 | 60 | 8 | 7 |

Table 38.4. Whiting in Division 6.a. Discards-at-age (thousands). Previous discard estimates (ICES, WGCSE 2011) for the years 1978-2003 were replaced by those estimated by Millar and Fryer (2005).

| Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1965 | 17205 | 4968 | 11437 | 531 | 14 | 2 | 0 |
| 1966 | 4322 | 8946 | 515 | 3317 | 79 | 3 | 0 |
| 1967 | 12237 | 20791 | 2674 | 84 | 629 | 12 | 1 |
| 1968 | 16394 | 12612 | 2137 | 377 | 13 | 82 | 3 |
| 1969 | 1983 | 20494 | 2093 | 292 | 51 | 2 | 26 |
| 1970 | 1776 | 6704 | 7494 | 382 | 33 | 4 | 0 |
| 1971 | 5505 | 6719 | 969 | 3906 | 57 | 4 | 1 |
| 1972 | 39192 | 8930 | 850 | 152 | 610 | 14 | 1 |
| 1973 | 30521 | 26995 | 1225 | 147 | 14 | 77 | 2 |
| 1974 | 23101 | 40590 | 2362 | 123 | 7 | 1 | 7 |
| 1975 | 37295 | 13541 | 8485 | 310 | 12 | 1 | 0 |
| 1976 | 24891 | 35812 | 3360 | 1940 | 63 | 1 | 0 |
| 1977 | 48148 | 8675 | 5432 | 301 | 212 | 5 | 0 |
| 1978 | 17886 | 12512 | 501 | 194 | 0 | 40 | 0 |
| 1979 | 2581 | 12099 | 1113 | 264 | 34 | 0 | 0 |
| 1980 | 2725 | 4889 | 2003 | 366 | 86 | 12 | 0 |
| 1981 | 1128 | 10415 | 1397 | 201 | 27 | 12 | 0 |
| 1982 | 19511 | 3421 | 12683 | 1197 | 187 | 4 | 0 |
| 1983 | 21690 | 6748 | 2909 | 5372 | 158 | 8 | 0 |
| 1984 | 34330 | 2400 | 909 | 371 | 811 | 73 | 1 |
| 1985 | 17615 | 9858 | 3273 | 672 | 205 | 363 | 40 |
| 1986 | 6159 | 9823 | 1962 | 185 | 1 | 0 | 10 |
| 1987 | 97611 | 17427 | 1763 | 154 | 0 | 0 | 0 |
| 1988 | 28057 | 38019 | 2239 | 467 | 11 | 0 | 0 |
| 1989 | 31079 | 5598 | 8570 | 223 | 13 | 5 | 0 |
| 1990 | 20952 | 11176 | 71 | 23 | 3 | 0 | 0 |
| 1991 | 23211 | 7540 | 7355 | 266 | 236 | 56 | 0 |


| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1992 | 50665 | 16729 | 2810 | 954 | 0 | 0 | 0 |
| 1993 | 14057 | 11139 | 2903 | 588 | 431 | 0 | 1 |
| 1994 | 12700 | 6859 | 3872 | 1152 | 189 | 150 | 4 |
| 1995 | 21974 | 21786 | 3416 | 484 | 7 | 1 | 1 |
| 1996 | 33621 | 18625 | 5086 | 1535 | 13 | 1 | 20 |
| 1997 | 22422 | 9632 | 3806 | 540 | 71 | 2 | 1 |
| 1998 | 53742 | 16058 | 3553 | 847 | 177 | 31 | 8 |
| 1999 | 7928 | 17097 | 1402 | 503 | 275 | 44 | 0 |
| 2000 | 158913 | 5254 | 2238 | 154 | 16 | 41 | 0 |
| 2001 | 5666 | 23084 | 715 | 172 | 0 | 0 | 0 |
| 2002 | 11055 | 8531 | 2428 | 415 | 175 | 9 | 3 |
| 2003 | 3770 | 1416 | 334 | 374 | 32 | 9 | 4 |
| 2004 | 14667 | 3557 | 536 | 305 | 107 | 4 | 2 |
| 2005 | 2923 | 1578 | 534 | 37 | 19 | 7 | 4 |
| 2006 | 9784 | 852 | 1000 | 256 | 36 | 11 | 2 |
| 2007 | 995 | 1077 | 308 | 64 | 4 | 3 | 0 |
| 2008 | 806 | 638 | 142 | 162 | 51 | 41 | 0 |
| 2009 | 6926 | 112 | 72 | 49 | 16 | 3 | 0 |
| 2010 | 16005 | 1427 | 245 | 42 | 61 | 6 | 1 |
| 2011 | 2697 | 1410 | 172 | 12 | 3 | 0 | 0 |
| 2012 | 7837 | 434 | 576 | 106 | 21 | 2 | 0 |
| 2013 | 13156 | 1338 | 159 | 252 | 12 | 3 | 2 |
| 2014 | 10618 | 44 | 71 | 35 | 36 | 10 | 3 |
| 2015 | 7550 | 866 | 284 | 119 | 20 | 17 | 0 |
| 2016 | 4640 | 1736 | 261 | 15 | 11 | 4 | 1 |
| 2017 | 9007 | 916 | 1005 | 68 | 189 | 2 | 0 |
| 2018 | 2673 | 846 | 316 | 196 | 22 | 5 | 0 |

Table 38.5. Whiting in Division 6.a. Total catch-at-age (thousands).

|  | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1965 | 24143 | 11054 | 54967 | 5334 | 402 | 105 | 22 |
| 1966 | 6007 | 19490 | 2744 | 31502 | 1940 | 189 | 53 |
| 1967 | 17406 | 46814 | 13293 | 781 | 15204 | 801 | 144 |
| 1968 | 23659 | 29096 | 11376 | 4034 | 337 | 5118 | 372 |
| 1969 | 2856 | 45668 | 10737 | 2858 | 1257 | 120 | 2358 |
| 1970 | 2506 | 13128 | 35559 | 3623 | 703 | 218 | 550 |
| 1971 | 7891 | 15336 | 5090 | 38690 | 1395 | 245 | 224 |
| 1972 | 55969 | 20958 | 4863 | 1514 | 15406 | 807 | 149 |
| 1973 | 44599 | 63137 | 6817 | 1608 | 371 | 4369 | 313 |
| 1974 | 32185 | 91625 | 12412 | 1289 | 188 | 53 | 856 |
| 1975 | 52213 | 30319 | 44804 | 3129 | 293 | 58 | 245 |
| 1976 | 33392 | 82233 | 19117 | 19363 | 1571 | 67 | 57 |
| 1977 | 64268 | 22051 | 30576 | 3428 | 4931 | 297 | 24 |
| 1978 | 35556 | 30687 | 7183 | 9594 | 941 | 1473 | 68 |
| 1979 | 8915 | 46320 | 14395 | 3671 | 3522 | 276 | 384 |
| 1980 | 14375 | 16267 | 16863 | 4521 | 1330 | 1097 | 190 |
| 1981 | 4721 | 34810 | 12694 | 4812 | 1545 | 464 | 201 |
| 1982 | 22502 | 9204 | 41777 | 8018 | 2230 | 807 | 348 |
| 1983 | 25108 | 13842 | 10949 | 28129 | 6228 | 1447 | 540 |
| 1984 | 41539 | 15165 | 9130 | 4758 | 15636 | 2026 | 859 |
| 1985 | 21754 | 29378 | 11847 | 4023 | 2202 | 5127 | 862 |
| 1986 | 8833 | 24647 | 11732 | 2838 | 533 | 291 | 539 |
| 1987 | 104041 | 31362 | 15751 | 5596 | 837 | 330 | 259 |
| 1988 | 29899 | 58606 | 11877 | 6635 | 1960 | 290 | 207 |
| 1989 | 33608 | 11485 | 20459 | 4990 | 1279 | 473 | 71 |
| 1990 | 24155 | 19204 | 2464 | 4032 | 1329 | 204 | 37 |
| 1991 | 26505 | 16366 | 17401 | 1474 | 1627 | 342 | 51 |


|  | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1992 | 53360 | 26169 | 7283 | 5736 | 396 | 373 | 106 |
| 1993 | 15108 | 21318 | 9196 | 3261 | 3169 | 163 | 148 |
| 1994 | 13609 | 11748 | 13030 | 4759 | 901 | 865 | 73 |
| 1995 | 22189 | 26108 | 9932 | 6138 | 1404 | 377 | 283 |
| 1996 | 34611 | 24035 | 12761 | 6587 | 2474 | 584 | 177 |
| 1997 | 23299 | 13290 | 12320 | 4856 | 1512 | 340 | 107 |
| 1998 | 54582 | 19562 | 7830 | 4545 | 1619 | 369 | 296 |
| 1999 | 8941 | 23228 | 5948 | 2543 | 2049 | 399 | 112 |
| 2000 | 159397 | 8206 | 6449 | 1724 | 501 | 369 | 89 |
| 2001 | 6127 | 26355 | 3345 | 1739 | 401 | 131 | 16 |
| 2002 | 11117 | 10155 | 5446 | 1214 | 402 | 32 | 16 |
| 2003 | 3940 | 2126 | 1445 | 2047 | 379 | 120 | 6 |
| 2004 | 14721 | 4281 | 1079 | 826 | 729 | 82 | 31 |
| 2005 | 2951 | 1854 | 989 | 177 | 118 | 52 | 11 |
| 2006 | 9866 | 991 | 1369 | 516 | 97 | 124 | 26 |
| 2007 | 1182 | 1245 | 563 | 390 | 136 | 29 | 50 |
| 2008 | 812 | 903 | 536 | 498 | 203 | 96 | 24 |
| 2009 | 6985 | 328 | 325 | 478 | 116 | 47 | 13 |
| 2010 | 16058 | 1521 | 399 | 161 | 187 | 30 | 32 |
| 2011 | 2697 | 1720 | 305 | 93 | 32 | 17 | 12 |
| 2012 | 7846 | 460 | 952 | 316 | 78 | 16 | 11 |
| 2013 | 13177 | 1388 | 243 | 529 | 79 | 21 | 8 |
| 2014 | 10630 | 75 | 202 | 137 | 136 | 33 | 14 |
| 2015 | 7561 | 949 | 345 | 283 | 88 | 84 | 25 |
| 2016 | 4641 | 1809 | 427 | 90 | 85 | 21 | 29 |
| 2017 | 9023 | 951 | 1172 | 139 | 238 | 39 | 10 |
| 2018 | 2674 | 917 | 404 | 395 | 82 | 13 | 7 |

Table 38.6. Whiting in Division 6.a. Landings weight-at-age (kg).

|  | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1965 | 0.218 | 0.249 | 0.308 | 0.452 | 1.208 | 0.72 | 0.778 |
| 1966 | 0.238 | 0.243 | 0.325 | 0.374 | 0.61 | 0.72 | 0.828 |
| 1967 | 0.204 | 0.24 | 0.319 | 0.424 | 0.412 | 0.639 | 0.821 |
| 1968 | 0.206 | 0.263 | 0.366 | 0.444 | 0.554 | 0.538 | 0.735 |
| 1969 | 0.178 | 0.223 | 0.335 | 0.5 | 0.57 | 0.649 | 0.63 |
| 1970 | 0.205 | 0.203 | 0.274 | 0.382 | 0.519 | 0.619 | 0.683 |
| 1971 | 0.209 | 0.247 | 0.276 | 0.316 | 0.426 | 0.551 | 0.712 |
| 1972 | 0.211 | 0.258 | 0.345 | 0.368 | 0.426 | 0.494 | 0.638 |
| 1973 | 0.196 | 0.235 | 0.362 | 0.479 | 0.485 | 0.532 | 0.666 |
| 1974 | 0.193 | 0.215 | 0.317 | 0.444 | 0.591 | 0.641 | 0.584 |
| 1975 | 0.209 | 0.245 | 0.305 | 0.471 | 0.651 | 0.615 | 0.717 |
| 1976 | 0.201 | 0.242 | 0.309 | 0.361 | 0.497 | 0.687 | 0.856 |
| 1977 | 0.2 | 0.244 | 0.296 | 0.392 | 0.431 | 0.629 | 0.819 |
| 1978 | 0.199 | 0.235 | 0.286 | 0.389 | 0.516 | 0.549 | 0.612 |
| 1979 | 0.218 | 0.232 | 0.306 | 0.404 | 0.536 | 0.678 | 0.693 |
| 1980 | 0.172 | 0.242 | 0.33 | 0.42 | 0.492 | 0.595 | 0.817 |
| 1981 | 0.192 | 0.228 | 0.289 | 0.382 | 0.409 | 0.409 | 0.547 |
| 1982 | 0.184 | 0.22 | 0.276 | 0.352 | 0.505 | 0.513 | 0.526 |
| 1983 | 0.216 | 0.249 | 0.28 | 0.34 | 0.409 | 0.494 | 0.51 |
| 1984 | 0.216 | 0.259 | 0.313 | 0.371 | 0.412 | 0.458 | 0.458 |
| 1985 | 0.185 | 0.238 | 0.306 | 0.402 | 0.43 | 0.461 | 0.538 |
| 1986 | 0.174 | 0.236 | 0.294 | 0.365 | 0.468 | 0.482 | 0.499 |
| 1987 | 0.188 | 0.237 | 0.304 | 0.373 | 0.511 | 0.52 | 0.576 |
| 1988 | 0.176 | 0.215 | 0.301 | 0.4 | 0.483 | 0.567 | 0.6 |
| 1989 | 0.171 | 0.22 | 0.279 | 0.348 | 0.459 | 0.425 | 0.555 |
| 1990 | 0.225 | 0.251 | 0.324 | 0.359 | 0.417 | 0.582 | 0.543 |
| 1991 | 0.199 | 0.22 | 0.291 | 0.354 | 0.391 | 0.442 | 0.761 |


| Year | Age <br> 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.193 | 0.23 | 0.288 | 0.349 | 0.388 | 0.397 | 0.51 |
| 1993 | 0.186 | 0.242 | 0.314 | 0.361 | 0.412 | 0.452 | 0.474 |
| 1994 | 0.161 | 0.217 | 0.29 | 0.371 | 0.451 | 0.482 | 0.483 |
| 1995 | 0.19 | 0.225 | 0.296 | 0.381 | 0.469 | 0.473 | 0.528 |
| 1996 | 0.195 | 0.245 | 0.288 | 0.365 | 0.483 | 0.526 | 0.569 |
| 1997 | 0.198 | 0.245 | 0.297 | 0.384 | 0.522 | 0.629 | 0.661 |
| 1998 | 0.215 | 0.236 | 0.301 | 0.364 | 0.438 | 0.5 | 0.646 |
| 1999 | 0.181 | 0.225 | 0.28 | 0.365 | 0.44 | 0.524 | 0.594 |
| 2000 | 0.205 | 0.241 | 0.298 | 0.336 | 0.419 | 0.488 | 0.617 |
| 2001 | 0.173 | 0.234 | 0.303 | 0.37 | 0.395 | 0.376 | 0.595 |
| 2002 | 0.213 | 0.257 | 0.304 | 0.363 | 0.464 | 0.65 | 0.707 |
| 2003 | 0.228 | 0.264 | 0.309 | 0.362 | 0.374 | 0.436 | 0.717 |
| 2004 | 0.193 | 0.251 | 0.295 | 0.345 | 0.382 | 0.403 | 0.342 |
| 2005 | 0.189 | 0.261 | 0.313 | 0.378 | 0.44 | 0.482 | 0.356 |
| 2006 | 0.221 | 0.292 | 0.319 | 0.394 | 0.455 | 0.528 | 0.567 |
| 2007 | 0.215 | 0.280 | 0.349 | 0.418 | 0.498 | 0.598 | 0.660 |
| 2008 | 0.274 | 0.245 | 0.322 | 0.384 | 0.514 | 0.530 | 0.653 |
| 2009 | 0.328 | 0.347 | 0.437 | 0.479 | 0.470 | 0.519 | 0.595 |
| 2010 | 0.288 | 0.402 | 0.456 | 0.567 | 0.652 | 0.619 | 0.613 |
| 2011 | 0.210 | 0.327 | 0.405 | 0.523 | 0.613 | 0.570 | 0.393 |
| 2012 | 0.295 | 0.304 | 0.387 | 0.508 | 0.615 | 0.705 | 0.493 |
| 2013 | 0.191 | 0.277 | 0.354 | 0.442 | 0.541 | 0.631 | 0.729 |
| 2014 | 0.243 | 0.271 | 0.374 | 0.463 | 0.544 | 0.659 | 0.699 |
| 2015 | 0.290 | 0.356 | 0.444 | 0.467 | 0.513 | 0.601 | 0.624 |
| 2016 | 0.272 | 0.402 | 0.520 | 0.543 | 0.614 | 0.700 | 0.693 |
| 2017 | 0.341 | 0.353 | 0.418 | 0.544 | 0.500 | 0.507 | 0.677 |
| 2018 | 0.177 | 0.408 | 0.394 | 0.433 | 0.516 | 0.468 | 0.574 |

Table 38.7. Whiting in Division 6.a. Discards weight-at-age (kg).

| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1965 | 0.122 | 0.177 | 0.213 | 0.249 | 0.287 | 0.303 | 0.287 |
| 1966 | 0.122 | 0.178 | 0.212 | 0.248 | 0.29 | 0.297 | 0.286 |
| 1967 | 0.122 | 0.178 | 0.213 | 0.248 | 0.29 | 0.295 | 0.289 |
| 1968 | 0.128 | 0.179 | 0.213 | 0.249 | 0.291 | 0.298 | 0.287 |
| 1969 | 0.121 | 0.178 | 0.214 | 0.249 | 0.29 | 0.295 | 0.285 |
| 1970 | 0.121 | 0.175 | 0.213 | 0.249 | 0.29 | 0.299 | 0.284 |
| 1971 | 0.12 | 0.177 | 0.211 | 0.248 | 0.29 | 0.299 | 0.284 |
| 1972 | 0.121 | 0.177 | 0.213 | 0.248 | 0.289 | 0.301 | 0.281 |
| 1973 | 0.123 | 0.176 | 0.215 | 0.252 | 0.288 | 0.301 | 0.285 |
| 1974 | 0.119 | 0.177 | 0.214 | 0.25 | 0.285 | 0.299 | 0.288 |
| 1975 | 0.119 | 0.176 | 0.213 | 0.25 | 0.286 | 0.301 | 0.278 |
| 1976 | 0.116 | 0.177 | 0.213 | 0.249 | 0.288 | 0.3 | 0.28 |
| 1977 | 0.118 | 0.177 | 0.214 | 0.249 | 0.289 | 0.299 | 0.282 |
| 1978 | 0.135 | 0.167 | 0.199 | 0.288 | 0.32 | 0.238 | - |
| 1979 | 0.173 | 0.188 | 0.208 | 0.215 | 0.281 | - | - |
| 1980 | 0.14 | 0.179 | 0.208 | 0.22 | 0.271 | 0.386 | - |
| 1981 | 0.108 | 0.16 | 0.195 | 0.298 | 0.286 | 0.295 | - |
| 1982 | 0.096 | 0.18 | 0.209 | 0.243 | 0.283 | 0.44 | - |
| 1983 | 0.141 | 0.186 | 0.228 | 0.237 | 0.267 | 0.267 | - |
| 1984 | 0.087 | 0.199 | 0.246 | 0.26 | 0.259 | 0.303 | 0.227 |
| 1985 | 0.102 | 0.191 | 0.237 | 0.286 | 0.326 | 0.312 | 0.316 |
| 1986 | 0.092 | 0.17 | 0.196 | 0.245 | 0.258 | 0.33 | 0.263 |
| 1987 | 0.085 | 0.182 | 0.233 | 0.249 | 0.225 | - | - |
| 1988 | 0.076 | 0.143 | 0.203 | 0.227 | 0.262 | - | - |
| 1989 | 0.099 | 0.177 | 0.205 | 0.209 | 0.294 | 0.305 | - |
| 1990 | 0.124 | 0.171 | 0.214 | 0.219 | 0.237 | 0.264 | - |
| 1991 | 0.085 | 0.169 | 0.205 | 0.223 | 0.226 | 0.281 | - |


| Year | Age <br> 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.109 | 0.173 | 0.219 | 0.227 | - | - | - |
| 1993 | 0.118 | 0.197 | 0.225 | 0.242 | 0.256 | - | 0.436 |
| 1994 | 0.087 | 0.157 | 0.22 | 0.283 | 0.297 | 0.253 | 0.299 |
| 1995 | 0.075 | 0.154 | 0.189 | 0.246 | 0.278 | 0.597 | 0.493 |
| 1996 | 0.095 | 0.18 | 0.203 | 0.229 | 0.302 | 0.421 | 0.26 |
| 1997 | 0.112 | 0.182 | 0.221 | 0.235 | 0.243 | 0.422 | 0.819 |
| 1998 | 0.098 | 0.179 | 0.225 | 0.254 | 0.282 | 0.264 | 0.245 |
| 1999 | 0.077 | 0.168 | 0.217 | 0.205 | 0.266 | 0.268 | - |
| 2000 | 0.075 | 0.164 | 0.203 | 0.233 | 0.282 | 0.25 | - |
| 2001 | 0.094 | 0.154 | 0.196 | 0.203 | 0.381 | - | - |
| 2002 | 0.073 | 0.162 | 0.212 | 0.245 | 0.24 | 0.295 | 0.276 |
| 2003 | 0.077 | 0.177 | 0.231 | 0.242 | 0.213 | 0.3 | 0.278 |
| 2004 | 0.086 | 0.186 | 0.236 | 0.246 | 0.304 | 0.349 | 0.314 |
| 2005 | 0.088 | 0.149 | 0.223 | 0.214 | 0.315 | 0.292 | 0.373 |
| 2006 | 0.046 | 0.197 | 0.235 | 0.295 | 0.322 | 0.518 | 0.362 |
| 2007 | 0.059 | 0.159 | 0.225 | 0.226 | 0.334 | 0.794 | 0.266 |
| 2008 | 0.075 | 0.211 | 0.286 | 0.301 | 0.397 | 0.222 | 0.304 |
| 2009 | 0.051 | 0.288 | 0.227 | 0.262 | 0.248 | 0.253 | - |
| 2010 | 0.038 | 0.124 | 0.269 | 0.375 | 0.376 | 0.401 | 0.964 |
| 2011 | 0.030 | 0.141 | 0.321 | 0.266 | 0.221 | - | - |
| 2012 | 0.057 | 0.151 | 0.292 | 0.355 | 0.349 | 0.414 | 0.907 |
| 2013 | 0.041 | 0.208 | 0.238 | 0.355 | 0.377 | 0.297 | 0.371 |
| 2014 | 0.049 | 0.168 | 0.279 | 0.364 | 0.442 | 0.441 | 0.791 |
| 2015 | 0.074 | 0.181 | 0.226 | 0.349 | 0.322 | 0.440 | - |
| 2016 | 0.073 | 0.216 | 0.282 | 0.292 | 0.310 | 0.261 | 0.384 |
| 2017 | 0.065 | 0.197 | 0.348 | 0.411 | 0.328 | 0.881 | - |
| 2018 | 0.066 | 0.178 | 0.267 | 0.337 | 0.404 | 1.107 | - |

Table 38.8. Whiting in Division 6.a. Total catch weight-at-age (kg).

|  | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1965 | 0.150 | 0.217 | 0.288 | 0.432 | 1.176 | 0.712 | 0.778 |
| 1966 | 0.155 | 0.213 | 0.304 | 0.361 | 0.597 | 0.713 | 0.812 |
| 1967 | 0.146 | 0.212 | 0.298 | 0.405 | 0.407 | 0.634 | 0.817 |
| 1968 | 0.152 | 0.227 | 0.337 | 0.426 | 0.544 | 0.534 | 0.729 |
| 1969 | 0.138 | 0.203 | 0.311 | 0.474 | 0.559 | 0.643 | 0.626 |
| 1970 | 0.145 | 0.189 | 0.261 | 0.368 | 0.508 | 0.613 | 0.683 |
| 1971 | 0.147 | 0.216 | 0.264 | 0.309 | 0.420 | 0.545 | 0.710 |
| 1972 | 0.148 | 0.223 | 0.322 | 0.356 | 0.421 | 0.491 | 0.636 |
| 1973 | 0.146 | 0.210 | 0.336 | 0.458 | 0.478 | 0.528 | 0.661 |
| 1974 | 0.140 | 0.198 | 0.297 | 0.425 | 0.576 | 0.635 | 0.582 |
| 1975 | 0.145 | 0.214 | 0.288 | 0.449 | 0.636 | 0.610 | 0.717 |
| 1976 | 0.138 | 0.214 | 0.292 | 0.350 | 0.489 | 0.681 | 0.856 |
| 1977 | 0.139 | 0.218 | 0.281 | 0.379 | 0.425 | 0.623 | 0.819 |
| 1978 | 0.160 | 0.210 | 0.276 | 0.387 | 0.516 | 0.545 | 0.612 |
| 1979 | 0.202 | 0.222 | 0.295 | 0.378 | 0.530 | 0.678 | 0.693 |
| 1980 | 0.167 | 0.220 | 0.308 | 0.393 | 0.467 | 0.594 | 0.817 |
| 1981 | 0.173 | 0.196 | 0.271 | 0.379 | 0.402 | 0.408 | 0.547 |
| 1982 | 0.109 | 0.202 | 0.252 | 0.336 | 0.499 | 0.513 | 0.526 |
| 1983 | 0.155 | 0.215 | 0.270 | 0.324 | 0.405 | 0.479 | 0.510 |
| 1984 | 0.099 | 0.245 | 0.305 | 0.358 | 0.397 | 0.454 | 0.456 |
| 1985 | 0.107 | 0.216 | 0.288 | 0.383 | 0.427 | 0.448 | 0.537 |
| 1986 | 0.109 | 0.198 | 0.274 | 0.360 | 0.465 | 0.481 | 0.474 |
| 1987 | 0.097 | 0.210 | 0.297 | 0.369 | 0.510 | 0.520 | 0.576 |
| 1988 | 0.080 | 0.164 | 0.281 | 0.392 | 0.477 | 0.567 | 0.600 |
| 1989 | 0.108 | 0.204 | 0.255 | 0.337 | 0.446 | 0.422 | 0.555 |
| 1990 | 0.140 | 0.217 | 0.295 | 0.342 | 0.405 | 0.575 | 0.543 |
| 1991 | 0.096 | 0.207 | 0.265 | 0.338 | 0.376 | 0.424 | 0.761 |


| Year | Age <br> 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.114 | 0.195 | 0.265 | 0.329 | 0.388 | 0.397 | 0.510 |
| 1993 | 0.123 | 0.211 | 0.271 | 0.331 | 0.361 | 0.452 | 0.473 |
| 1994 | 0.089 | 0.170 | 0.258 | 0.344 | 0.419 | 0.448 | 0.473 |
| 1995 | 0.076 | 0.166 | 0.235 | 0.361 | 0.440 | 0.472 | 0.526 |
| 1996 | 0.098 | 0.198 | 0.257 | 0.336 | 0.482 | 0.526 | 0.537 |
| 1997 | 0.116 | 0.200 | 0.275 | 0.369 | 0.505 | 0.629 | 0.661 |
| 1998 | 0.101 | 0.197 | 0.274 | 0.341 | 0.420 | 0.469 | 0.573 |
| 1999 | 0.084 | 0.194 | 0.269 | 0.341 | 0.433 | 0.505 | 0.594 |
| 2000 | 0.076 | 0.199 | 0.277 | 0.329 | 0.415 | 0.477 | 0.617 |
| 2001 | 0.100 | 0.183 | 0.280 | 0.350 | 0.395 | 0.376 | 0.560 |
| 2002 | 0.074 | 0.194 | 0.270 | 0.346 | 0.385 | 0.541 | 0.728 |
| 2003 | 0.080 | 0.211 | 0.287 | 0.340 | 0.360 | 0.424 | 0.498 |
| 2004 | 0.086 | 0.197 | 0.266 | 0.308 | 0.371 | 0.400 | 0.340 |
| 2005 | 0.089 | 0.166 | 0.264 | 0.344 | 0.420 | 0.456 | 0.362 |
| 2006 | 0.047 | 0.210 | 0.258 | 0.345 | 0.406 | 0.527 | 0.551 |
| 2007 | 0.084 | 0.175 | 0.281 | 0.387 | 0.494 | 0.616 | 0.659 |
| 2008 | 0.076 | 0.221 | 0.312 | 0.357 | 0.484 | 0.397 | 0.649 |
| 2009 | 0.053 | 0.327 | 0.391 | 0.457 | 0.440 | 0.500 | 0.595 |
| 2010 | 0.038 | 0.141 | 0.341 | 0.517 | 0.562 | 0.573 | 0.622 |
| 2011 | 0.030 | 0.174 | 0.358 | 0.491 | 0.571 | 0.570 | 0.393 |
| 2012 | 0.058 | 0.160 | 0.329 | 0.456 | 0.543 | 0.673 | 0.497 |
| 2013 | 0.041 | 0.211 | 0.278 | 0.401 | 0.516 | 0.583 | 0.658 |
| 2014 | 0.050 | 0.210 | 0.341 | 0.438 | 0.517 | 0.593 | 0.720 |
| 2015 | 0.074 | 0.196 | 0.264 | 0.417 | 0.470 | 0.567 | 0.624 |
| 2016 | 0.073 | 0.224 | 0.374 | 0.500 | 0.573 | 0.612 | 0.680 |
| 2017 | 0.066 | 0.203 | 0.358 | 0.479 | 0.363 | 0.521 | 0.677 |
| 2018 | 0.066 | 0.195 | 0.294 | 0.385 | 0.486 | 0.729 | 0.574 |

Table 38.9. Whiting in Division 6.a. Survey data made available to the WG. Data used in the TSA run are highlighted in bold. For the Scottish surveys, numbers are standardised to catch-rate per ten hours. The Scottish surveys from 2011 have been conducted according to the new design and ground gear.

| Year | ScoGFS-WIBTS-Q1 - Scottish Groundfish Survey - numbers-at-age/10 h |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |  |
|  | (hours) | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1985 | 10 | 3140 | 1792 | 380 | 85 | 23 | 156 | 18 |
| 1986 | 10 | 1456 | 1525 | 403 | 68 | 10 | 9 | 10 |
| 1987 | 10 | 6938 | 1054 | 584 | 142 | 36 | 2 | 1 |
| 1988 | 10 | 567 | 3469 | 654 | 189 | 42 | 5 | 1 |
| 1989 | 10 | 910 | 505 | 586 | 237 | 48 | 3 | 0 |
| 1990 | 10 | 1818 | 571 | 122 | 216 | 61 | 4 | 1 |
| 1991 | 10 | 3203 | 276 | 299 | 22 | 39 | 9 | 1 |
| 1992 | 10 | 4777 | 1597 | 410 | 517 | 56 | 18 | 0 |
| 1993 | 10 | 5532 | 6829 | 644 | 91 | 30 | 11 | 2 |
| 1994 | 10 | 6614 | 2443 | 1487 | 174 | 56 | 15 | 6 |
| 1995 | 10 | 5598 | 2831 | 1160 | 370 | 70 | 17 | 32 |
| 1996 | 10 | 9385 | 2237 | 635 | 341 | 135 | 30 | 4 |
| 1997 | 10 | 5663 | 2444 | 1531 | 355 | 102 | 17 | 4 |
| 1998 | 10 | 9851 | 1352 | 294 | 195 | 50 | 14 | 1 |
| 1999 | 10 | 6125 | 4952 | 489 | 103 | 16 | 1 | 0 |
| 2000 | 10 | 12862 | 471 | 152 | 34 | 10 | 11 | 0 |
| 2001 | 10 | 4653 | 1955 | 242 | 41 | 8 | 1 | 1 |
| 2002 | 10 | 5542 | 1028 | 964 | 89 | 15 | 1 | 1 |
| 2003 | 10 | 6934 | 746 | 436 | 300 | 32 | 2 | 4 |
| 2004 | 10 | 5887 | 1566 | 189 | 131 | 44 | 9 | 1 |
| 2005 | 10 | 1308 | 723 | 183 | 35 | 8 | 11 | 2 |
| 2006 | 10 | 1441 | 466 | 282 | 77 | 0 | 3 | 1 |
| 2007 | 10 | 614 | 522 | 127 | 75 | 16 | 3 | 2 |
| 2008 | 10 | 593 | 127 | 77 | 26 | 8 | 3 | 0 |
| 2009 | 10 | 906 | 387 | 103 | 105 | 20 | 9 | 7 |
| 2010 | 10 | 3523 | 340 | 108 | 52 | 40 | 4 | 3 |

Table 38.9. Continued.

| Year | ScoGFS-WIBTS-Q4 - Scottish Groundfish Survey - numbers-at-age/10 h |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |  |  |
|  | (hours) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1996 | 10 | 5154 | 1908 | 1116 | 570 | 188 | 51 | 6 | 1 |
| 1997 | 10 | 8001 | 2869 | 951 | 323 | 160 | 46 | 12 | 1 |
| 1998 | 10 | 1852 | 2713 | 1125 | 150 | 100 | 20 | 1 | 0 |
| 1999 | 10 | 8203 | 2338 | 582 | 141 | 33 | 24 | 1 | 1 |
| 2000 | 10 | 4434 | 4056 | 789 | 160 | 9 | 7 | 1 | 0 |
| 2001 | 10 | 9615 | 1957 | 1420 | 155 | 40 | 12 | 2 | 0 |
| 2002 | 10 | 14658 | 1591 | 621 | 479 | 30 | 9 | 5 | 0 |
| 2003 | 10 | 9932 | 3446 | 567 | 338 | 83 | 27 | 4 | 0 |
| 2004 | 10 | 5923 | 1758 | 940 | 83 | 57 | 62 | 1 | 0 |
| 2005 | 10 | 2297 | 308 | 318 | 76 | 9 | 4 | 1 | 1 |
| 2006 | 10 | 415 | 296 | 140 | 101 | 35 | 8 | 3 | 0 |
| 2007 | 10 | 1894 | 434 | 326 | 99 | 83 | 48 | 1 | 0 |
| 2008 | 10 | 2297 | 208 | 78 | 110 | 28 | 24 | 4 | 0 |
| 2009 | 10 | 4833 | 236 | 178 | 50 | 58 | 12 | 6 | 6 |


| Year | IGFS-WIBTS-Q4 - Irish groundfish survey - numbers-at-age/10 h |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age |  |  |  |  |  |  |
|  | (hours) | 0 | 1 | 2 | 3 | 4 | 5 | 6 |
| 2003 | 10 | 586 | 6860 | 1541 | 273 | 154 | 54 | 1 |
| 2004 | 10 | 3462 | 1557 | 656 | 52 | 18 | 8 | 1 |
| 2005 | 10 | 569 | 1393 | 704 | 57 | 3 | 3 | 0 |
| 2006 | 10 | 39 | 419 | 366 | 85 | 11 | 1 | 0 |
| 2007 | 10 | 70 | 1018 | 1217 | 369 | 87 | 129 | 62 |
| 2008 | 10 | 13 | 2295 | 702 | 303 | 128 | 65 | 19 |
| 2009 | 10 | 7361 | 623 | 431 | 141 | 29 | 9 | 18 |
| 2010 | 10 | 50 | 4565 | 702 | 178 | 56 | 30 | 7 |
| 2011 | 10 | 211 | 2074 | 2817 | 318 | 135 | 32 | 33 |
| 2012 | 10 | 129 | 3226 | 499 | 970 | 276 | 24 | 11 |
| 2013 | 10 | 11247 | 494 | 1865 | 498 | 555 | 65 | 6 |
| 2014 | 10 | 14934 | 7930 | 1300 | 2618 | 300 | 356 | 30 |
| 2015 | 10 | 1862 | 15267 | 3237 | 794 | 400 | 81 | 54 |
| 2016 | 10 | 6404 | 5918 | 8840 | 1387 | 234 | 290 | 92 |
| 2017 | 10 | 252 | 1969 | 1414 | 1873 | 331 | 39 | 45 |
| 2018 | 10 | 8451 | 2357 | 2860 | 1853 | 712 | 42 | 0 |

Table 38.9. Continued.

| UK-SCOWCGFS-Q1 - Scottish Groundfish Survey - numbers-at-age/10 h |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Effort | Age |  |  |  |  |  |  |  |
|  | (hours) | 1 | 2 | 3 | 4 |  | 5 | 6 | 7 |
| 2011 | 10 | 222 | 1884 | 397 | 64 |  | 37 | 45 | 12 |
| 2012 | 10 | 3441 | 293 | 738 | 72 |  | 14 | 5 | 7 |
| 2013 | 10 | 552 | 1031 | 302 | 463 |  | 61 | 7 | 3 |
| 2014 | 10 | 5805 | 125 | 246 | 110 |  | 74 | 7 | 1 |
| 2015 | 10 | 2545 | 760 | 285 | 259 |  | 65 | 58 | 8 |
| 2016 | 10 | 3226 | 3485 | 576 | 148 |  | 84 | 42 | 25 |
| 2017 | 10 | 4970 | 1981 | 1707 | 203 |  | 49 | 32 | 5 |
| 2018 | 10 | 1960 | 1827 | 1069 | 1142 |  | 132 | 14 | 2 |
| 2019 | 10 | 3231 | 666 | 577 | 191 |  | 99 | 25 | 0 |
| UK-SCOWCGFS-Q4 - Scottish Groundfish Survey - numbers-at-age/10 h |  |  |  |  |  |  |  |  |  |
| Year | Effort | Age |  |  |  |  |  |  |  |
|  | (hours) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 2011 | 10 | 3644 | 119 | 2096 | 109 | 30 | 14 | 10 | 1 |
| 2012 | 10 | 748 | 964 | 426 | 658 | 110 | 19 | 2 | 11 |
| 2013 | 10 | 1732 | 125 | 309 | 110 | 159 | 27 | 2 | 0 |
| 2014 | 10 | 11569 | 1518 | 346 | 168 | 82 | 55 | 31 | 0 |
| 2015 | 10 | 4263 | 2794 | 727 | 115 | 91 | 20 | 27 | 1 |
| 2016 | 10 | 5262 | 2415 | 2300 | 259 | 83 | 115 | 29 | 13 |
| 2017 | 10 | 3306 | 2943 | 4139 | 1167 | 177 | 2 | 12 | 2 |
| 2018 | 10 | 6442 | 503 | 552 | 284 | 220 | 33 | 1 | 5 |

Table 38.10. Whiting in Division 6.a. TSA parameter settings for the assessment run.

| Parameter | Setting | Justification |
| :---: | :---: | :---: |
| Age of full selection | $a_{m}=4$ | Based on inspection of previous XSA and TSA runs. |
| Multipliers on variance matrices of measurements | $B_{\text {landings }}(a)=2$ for ages $1,7+$ <br> $B_{\text {discards }}(a)=2$ for age 5 <br> $B_{\text {ScoGFs-WIBTS-Q4 }}(a)=2$ for age 6 | Allows extra measurement variability for poorly-sampled ages. |
| Multipliers on variances for fishing mortality estimates | $H(1)=2$ | Allows for more variable fishing mortalities for age 1 fish. |
| Down-weighting of particular datapoints | Discards: <br> cvmult = 3 for age 1 in 1981, age 3 in 1991, age 1 in 2000, age 1 in 2013, age 5 in 2017 <br> Surveys: <br> ScoGFS-WIBTS-Q1 <br> cvmult = 3 for age 5 in 1992, age 2 in 1993, age 1 in 2000 and age 2 in 2000 <br> cvmult $=5$ for age 4 in 1992 <br> ScoGFS-WIBTS-Q4 <br> cvmult $=3$ for age 4 in 2007 and for age 5 in 2007 | Large values indicated by exploratory prediction error plots. |
| Discards | Discards are allowed to evolve over time constrained by a trend. Ages 1 to 5 are modelled independently. |  |
| Recruitments | Modelled by a hockey-stick model, with numbers-at-age 1 assumed to be independent and normally distributed. To allow recruitment variability to increase with mean recruitment, a constant coefficient of variation is assumed. |  |

Table 38.11. Whiting in Division 6.a. TSA parameter estimates for final assessment presented this year.

| Parameter | Notation | Description | $\begin{aligned} & 2017 \\ & \text { WG } \end{aligned}$ | 2018 WG | 2019 WG |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Initial fishing mortality | $F(1,1981)$ | Fishing mortality-at-age $a$ in year $y$ | 0.09 | 0.0937 | 0.0945 |
|  | $F(2,1981)$ |  | 0.11 | 0.1054 | 0.1071 |
|  | $F(4,1981)$ |  | 0.32 | 0.3129 | 0.3207 |
| Fishing mortality standard deviations | $\sigma_{F}$ | Transitory changes in overall fishing mortality | 0.01 | 0.0319 | 0.0000 |
|  | $\sigma_{u}$ | Persistent changes in selection (age effect in F) | 0.09 | 0.0879 | 0.0884 |
|  | $\sigma_{v}$ | Transitory changes in the year effect in fishing mortality | 0.00 | 0.0008 | 0.0000 |
|  | $\sigma_{Y}$ | Persistent changes in the year effect in fishing mortality | 0.28 | 0.2654 | 0.2652 |
| Measurement CVs | $\mathrm{CV}_{\text {landings }}$ | CV of landings-at-age data | 0.16 | 0.1626 | 0.1688 |
|  | $\mathrm{CV}_{\text {discards }}$ | CV of discards-at-age data | 0.53 | 0.5276 | 0.5275 |
| Recruitment |  | Hockey-stick parameter <br> Recruitment value at change point | 29.4 | 28.9494 | 29.6289 |
|  |  | Hockey-stick parameter SSB at change point | 2.90 | 2.932 | 3.3912 |
|  | $\mathrm{CV}_{\text {rec }}$ | Coefficient of variation of recruitment data | 0.33 | 0.3407 | 0.3331 |
| Discards | $\sigma_{\text {ogit p }}$ | Transitory trends in discarding | 0.26 | 0.2621 | 0.2827 |
|  | $\sigma_{\text {persistent }}$ | Persistent trends in discarding | 0.22 | 0.2389 | 0.2275 |
| Survey selectivities (ScoGFS-WIBTS-Q1) | $\Phi(1)$ | Survey selectivity-at-age a | 1.01 | 0.9447 | 1.0934 |
|  | $\Phi(2)$ |  | 1.05 | 0.9838 | 1.1093 |
|  | $\Phi(3)$ |  | 0.89 | 0.8326 | 0.9403 |
|  | $\Phi(4)$ |  | 0.75 | 0.6929 | 0.7852 |
|  | $\Phi(5)$ |  | 0.60 | 0.554 | 0.6455 |
|  | $\Phi(6)$ |  | 0.53 | 0.4955 | 0.5628 |
|  | $\sigma_{\text {survey }}$ | Standard error of survey data | 0.45 | 0.4667 | 0.4549 |
|  | $\sigma_{\eta}$ |  | 0.10 | 0.1046 | 0.1056 |
| Survey catchability standard deviations | $\sigma_{\Omega}$ | Transitory changes in survey catchability | 0.22 | 0.2443 | 0.1916 |


| Parameter | Notation | Description | $\begin{aligned} & 2017 \\ & \text { WG } \end{aligned}$ | 2018 WG | 2019 WG |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\sigma_{\beta}$ | Persistent changes in survey catchability | 0.11 | 0.113 | 0.1084 |
| Survey selectivities (ScoGFS-WIBTS-Q4) | $\Phi(1)$ | Survey selectivity-at-age $a$ | 3.15 | 3.4002 | 3.5733 |
|  | $\Phi(2)$ |  | 3.00 | 3.2019 | 3.2021 |
|  | $\Phi(3)$ |  | 2.33 | 2.4699 | 2.4844 |
|  | $\Phi(4)$ |  | 1.99 | 2.1113 | 2.1317 |
|  | $\Phi(5)$ |  | 2.67 | 2.8075 | 2.8519 |
|  | $\Phi(6)$ |  | 0.48 | 0.4558 | 0.4902 |
|  | $\sigma_{\text {survey }}$ | Standard error of survey data | 0.20 | 0.1987 | 0.2146 |
|  | $\sigma_{\eta}$ |  | 0.19 | 0.2015 | 0.1993 |
| Survey catchability standard deviations | $\sigma_{\Omega}$ | Transitory changes in survey catchability | 0.00 | 0 | 0 |
|  | $\sigma_{\beta}$ | Persistent changes in survey catchability | 0.14 | 0.1413 | 0.1417 |
| Survey selectivities (IGFS-WI-BTS-Q4) | $\Phi(1)$ |  | 12.47 | 13.6417 | 13.3725 |
|  | $\Phi(2)$ |  | 11.60 | 12.7080 | 11.9858 |
|  | $\Phi(3)$ |  | 14.72 | 15.7623 | 15.1872 |
|  | $\Phi(4)$ |  | 10.26 | 11.3506 | 11.5093 |
|  | $\sigma_{\text {survey }}$ | Standard error of survey data | 0.27 | 0.2445 | 0.2769 |
|  | $\sigma_{\eta}$ |  | 0.47 | 0.4742 | 0.4365 |
| Survey catchability standard deviations | $\sigma_{\Omega}$ | Transitory changes in survey catchability | 0.08 | 0.1222 | 0.1578 |
|  | $\sigma_{\beta}$ | Persistent changes in survey catchability | 0.18 | 0.2016 | 0.1607 |
| Survey selectivities (UK-SCOWCGFS-Q1) | $\Phi(1)$ |  | 5.42 | 4.9149 | 6.0373 |
|  | $\Phi(2)$ |  | 6.40 | 5.8231 | 6.9723 |
|  | $\Phi(3)$ |  | 7.67 | 7.2411 | 8.0850 |
|  | $\Phi(4)$ |  | 6.32 | 6.6973 | 7.8329 |
|  | $\Phi(5)$ |  | 5.53 | 4.6753 | 5.4372 |
|  | $\Phi(6)$ |  | 6.30 | 4.8854 | 5.7624 |
|  | $\sigma_{\text {survey }}$ | Standard error of survey data | 0.36 | 0.3801 | 0.3763 |


| Parameter | Notation | Description | $\begin{aligned} & 2017 \\ & \text { WG } \end{aligned}$ | 2018 WG | 2019 WG |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\sigma_{\eta}$ |  | 0.16 | 0.2546 | 0.2062 |
| Survey catchability standard deviations | $\sigma_{\Omega}$ | Transitory changes in survey catchability | 0.01 | 0 | 0126 |
|  | $\sigma_{\beta}$ | Persistent changes in survey catchability | 0.06 | 0 | 0.1881 |
| Survey selectivities (UK-SCOWCGFS-Q4) | $\Phi(1)$ |  | 6.02 | 4.8981 | 4.7286 |
|  | $\Phi(2)$ |  | 10.36 | 13.8002 | 13.2288 |
|  | $\Phi(3)$ |  | 6.12 | 6.3429 | 5.7722 |
|  | $\Phi(4)$ |  | 7.89 | 6.7193 | 5.7311 |
|  | $\Phi(5)$ |  | 6.89 | 5.6839 | 4.4555 |
|  | $\Phi(6)$ |  | 8.09 | 6.3067 | 4.8454 |
|  | $\sigma_{\text {survey }}$ | Standard error of survey data | 0.26 | 0.3987 | 0.3842 |
|  | $\sigma_{\eta}$ |  | 0.01 | 0.0202 | 0.1695 |
| Survey catchability standard deviations | $\sigma_{\Omega}$ | Transitory changes in survey catchability | 0.13 | 0 | 0.1982 |
|  | $\sigma_{\beta}$ | Persistent changes in survey catchability | 0.01 | 0 | 0 |
| Misreporting |  | Transitory changes in misreporting | 0.00 | 0 | 0 |
|  |  | Persistent changes in misreporting | 0.16 | 0.1777 | 0.1915 |

Table 38.12. Whiting in Division 6.a. TSA population numbers-at-age (thousands).

| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 199423 | 470520 | 85528 | 22260 | 7075 | 2086 | 898 |
| 1982 | 165088 | 79380 | 218328 | 38485 | 9370 | 3059 | 1319 |
| 1983 | 197237 | 64095 | 35504 | 94594 | 15898 | 3988 | 1901 |
| 1984 | 325817 | 72150 | 24449 | 12346 | 31664 | 5399 | 2066 |
| 1985 | 310357 | 115409 | 24475 | 7282 | 3400 | 9323 | 2221 |
| 1986 | 291281 | 111599 | 37798 | 6003 | 1395 | 665 | 2546 |
| 1987 | 404905 | 110032 | 40829 | 12246 | 1492 | 352 | 828 |
| 1988 | 106310 | 143393 | 37376 | 12469 | 3282 | 337 | 241 |
| 1989 | 328200 | 34671 | 42998 | 10880 | 2557 | 618 | 41 |
| 1990 | 174358 | 120500 | 10835 | 11927 | 2636 | 476 | 65 |
| 1991 | 244996 | 63962 | 47051 | 3966 | 3581 | 796 | 115 |
| 1992 | 334144 | 91151 | 24074 | 17100 | 1339 | 1201 | 309 |
| 1993 | 262160 | 125125 | 34998 | 8976 | 5848 | 477 | 557 |
| 1994 | 274566 | 99114 | 48451 | 12757 | 2772 | 1866 | 335 |
| 1995 | 291943 | 105481 | 40474 | 18271 | 4104 | 914 | 740 |
| 1996 | 179537 | 112355 | 41028 | 14844 | 5235 | 1165 | 466 |
| 1997 | 163752 | 62677 | 41693 | 13599 | 3758 | 1293 | 401 |
| 1998 | 214790 | 54125 | 20662 | 13027 | 3216 | 887 | 404 |
| 1999 | 152810 | 66904 | 15536 | 5815 | 3016 | 708 | 284 |
| 2000 | 235620 | 43362 | 17598 | 3743 | 1091 | 588 | 194 |
| 2001 | 102705 | 68736 | 12529 | 4884 | 715 | 217 | 159 |
| 2002 | 40754 | 29785 | 20959 | 3726 | 977 | 138 | 76 |
| 2003 | 63845 | 9610 | 10253 | 7156 | 1034 | 281 | 63 |
| 2004 | 40215 | 16392 | 2788 | 3423 | 1706 | 259 | 88 |
| 2005 | 23483 | 10127 | 4842 | 823 | 855 | 402 | 86 |
| 2006 | 28198 | 7306 | 3884 | 1774 | 261 | 260 | 161 |
| 2007 | 15449 | 9126 | 2889 | 1521 | 593 | 89 | 147 |


| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 2008 | 16893 | 4747 | 3818 | 1224 | 549 | 225 | 91 |
| 2009 | 25105 | 5271 | 1824 | 1584 | 409 | 186 | 111 |
| 2010 | 59545 | 8460 | 2159 | 745 | 592 | 155 | 115 |
| 2011 | 19074 | 19951 | 3774 | 971 | 304 | 253 | 116 |
| 2012 | 38435 | 7121 | 9321 | 1862 | 466 | 147 | 183 |
| 2013 | 19782 | 14125 | 3427 | 4609 | 921 | 236 | 172 |
| 2014 | 45389 | 7587 | 6869 | 1757 | 2376 | 489 | 222 |
| 2015 | 105619 | 17416 | 3775 | 3557 | 938 | 1300 | 399 |
| 2016 | 66931 | 42094 | 8617 | 1959 | 1910 | 518 | 958 |
| 2017 | 79508 | 27279 | 21193 | 4535 | 1077 | 1077 | 856 |
| 2018 | 60000 | 32697 | 13833 | 11325 | 2528 | 617 | 1134 |
| 2019* | 140154 | 25191 | 16700 | 7476 | 6395 | 1467 | 1044 |
| 2020* | 175106 | 58676 | 12854 | 9014 | 4216 | 3704 | 1489 |
| GM(81-18) | 100335 | 37366 | 14907 | 5700 | 1839 | 598 | 300 |

[^21]Table 38.13. Whiting in Division 6.a. Standard errors on TSA population numbers-at-age (thousands).

| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 20088 | 34082 | 7457 | 1976 | 730 | 260 | 225 |
| 1982 | 17707 | 8104 | 16460 | 3675 | 931 | 355 | 171 |
| 1983 | 19493 | 7062 | 3947 | 8010 | 1733 | 468 | 227 |
| 1984 | 26806 | 7297 | 3208 | 1690 | 3302 | 795 | 296 |
| 1985 | 24040 | 9584 | 3087 | 1181 | 635 | 1438 | 451 |
| 1986 | 21630 | 8976 | 4060 | 1112 | 438 | 273 | 805 |
| 1987 | 34057 | 8438 | 3988 | 1641 | 447 | 192 | 452 |
| 1988 | 11437 | 12534 | 3262 | 1447 | 611 | 176 | 226 |
| 1989 | 22212 | 3972 | 4514 | 1249 | 553 | 249 | 131 |
| 1990 | 18899 | 8401 | 1314 | 1706 | 494 | 239 | 141 |
| 1991 | 23526 | 7103 | 3369 | 514 | 711 | 227 | 152 |
| 1992 | 29277 | 8870 | 2711 | 1330 | 202 | 295 | 140 |
| 1993 | 24459 | 11152 | 3527 | 1074 | 567 | 82 | 155 |
| 1994 | 27139 | 9737 | 4850 | 1591 | 456 | 281 | 93 |
| 1995 | 25465 | 11272 | 4595 | 2404 | 736 | 217 | 176 |
| 1996 | 20640 | 10435 | 4974 | 1996 | 945 | 315 | 165 |
| 1997 | 23369 | 7912 | 4117 | 1856 | 617 | 323 | 161 |
| 1998 | 33450 | 9064 | 3223 | 1598 | 613 | 227 | 162 |
| 1999 | 27455 | 12028 | 3405 | 1066 | 492 | 197 | 115 |
| 2000 | 38631 | 9291 | 3747 | 954 | 234 | 126 | 72 |
| 2001 | 15914 | 12718 | 2612 | 916 | 171 | 47 | 42 |
| 2002 | 8708 | 5336 | 3997 | 755 | 212 | 45 | 25 |
| 2003 | 11261 | 2600 | 1650 | 1368 | 198 | 61 | 20 |
| 2004 | 7129 | 3636 | 641 | 538 | 338 | 59 | 25 |
| 2005 | 3420 | 2067 | 856 | 178 | 114 | 94 | 25 |
| 2006 | 2578 | 787 | 452 | 199 | 31 | 26 | 32 |
| 2007 | 1852 | 802 | 268 | 168 | 67 | 12 | 21 |


| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 2008 | 1685 | 634 | 346 | 128 | 83 | 35 | 16 |
| 2009 | 2107 | 591 | 267 | 166 | 63 | 42 | 24 |
| 2010 | 5375 | 769 | 267 | 130 | 85 | 33 | 32 |
| 2011 | 1703 | 2019 | 357 | 136 | 67 | 47 | 33 |
| 2012 | 4454 | 675 | 1001 | 189 | 75 | 38 | 43 |
| 2013 | 2168 | 1795 | 337 | 543 | 107 | 42 | 42 |
| 2014 | 6285 | 896 | 914 | 185 | 310 | 63 | 45 |
| 2015 | 10118 | 2623 | 458 | 503 | 105 | 182 | 59 |
| 2016 | 9868 | 4295 | 1352 | 252 | 286 | 62 | 135 |
| 2017 | 19510 | 4188 | 2230 | 742 | 145 | 168 | 110 |
| 2018 | 23195 | 8332 | 2171 | 1240 | 426 | 86 | 155 |
| 2019* | 47647 | 9859 | 4299 | 1200 | 719 | 252 | 136 |
| 2020* | 69044 | 20190 | 5052 | 2350 | 689 | 429 | 219 |
| GM(81-18) | 11783 | 4484 | 1825 | 759 | 295 | 125 | 89 |

[^22]Table 38.14. Whiting in Division 6.a. TSA estimates for mortality-at-age.

|  | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 0.1029 | 0.1242 | 0.2178 | 0.3301 | 0.3301 | 0.3301 | 0.3301 |
| 1982 | 0.1152 | 0.1538 | 0.2590 | 0.3476 | 0.3475 | 0.3475 | 0.3475 |
| 1983 | 0.1775 | 0.2661 | 0.4310 | 0.5562 | 0.5562 | 0.5563 | 0.5562 |
| 1984 | 0.2193 | 0.3761 | 0.5461 | 0.6891 | 0.6889 | 0.6893 | 0.6888 |
| 1985 | 0.2324 | 0.4417 | 0.6223 | 0.7929 | 0.7929 | 0.7930 | 0.7931 |
| 1986 | 0.1815 | 0.3631 | 0.4920 | 0.5964 | 0.5962 | 0.5964 | 0.5963 |
| 1987 | 0.2230 | 0.4442 | 0.5932 | 0.7116 | 0.7114 | 0.7115 | 0.7114 |
| 1988 | 0.2560 | 0.5176 | 0.6456 | 0.8705 | 0.8707 | 0.8708 | 0.8708 |
| 1989 | 0.2273 | 0.4475 | 0.5950 | 0.7675 | 0.7675 | 0.7675 | 0.7675 |
| 1990 | 0.1728 | 0.3108 | 0.4285 | 0.5675 | 0.5675 | 0.5676 | 0.5676 |
| 1991 | 0.1761 | 0.3340 | 0.4379 | 0.5712 | 0.5712 | 0.5712 | 0.5712 |
| 1992 | 0.1688 | 0.3209 | 0.4310 | 0.5578 | 0.5577 | 0.5577 | 0.5577 |
| 1993 | 0.1706 | 0.3118 | 0.4371 | 0.6404 | 0.6404 | 0.6403 | 0.6404 |
| 1994 | 0.1556 | 0.2693 | 0.3926 | 0.5868 | 0.5873 | 0.5872 | 0.5871 |
| 1995 | 0.1799 | 0.3035 | 0.4256 | 0.6599 | 0.6599 | 0.6598 | 0.6602 |
| 1996 | 0.2412 | 0.3815 | 0.5255 | 0.8079 | 0.8077 | 0.8080 | 0.8078 |
| 1997 | 0.2843 | 0.4474 | 0.5939 | 0.8548 | 0.8554 | 0.8547 | 0.8551 |
| 1998 | 0.3331 | 0.5188 | 0.6663 | 0.9289 | 0.9291 | 0.9288 | 0.9288 |
| 1999 | 0.4048 | 0.6297 | 0.7833 | 1.1296 | 1.1297 | 1.1298 | 1.1298 |
| 2000 | 0.4010 | 0.5796 | 0.7259 | 1.1352 | 1.1352 | 1.1352 | 1.1352 |
| 2001 | 0.3796 | 0.5119 | 0.6260 | 1.0311 | 1.0311 | 1.0309 | 1.0311 |
| 2002 | 0.3026 | 0.3837 | 0.4613 | 0.7435 | 0.7436 | 0.7434 | 0.7434 |
| 2003 | 0.3456 | 0.3992 | 0.4660 | 0.8207 | 0.8211 | 0.8206 | 0.8211 |
| 2004 | 0.4022 | 0.4173 | 0.5115 | 0.8277 | 0.8272 | 0.8280 | 0.8274 |
| 2005 | 0.3765 | 0.3495 | 0.4461 | 0.6643 | 0.6646 | 0.6658 | 0.6644 |
| 2006 | 0.3620 | 0.2862 | 0.3713 | 0.5733 | 0.5729 | 0.5726 | 0.5731 |
| 2007 | 0.3307 | 0.2279 | 0.2806 | 0.4668 | 0.4669 | 0.4660 | 0.4663 |


| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 2008 | 0.3836 | 0.2649 | 0.3043 | 0.5300 | 0.5297 | 0.5295 | 0.5305 |
| 2009 | 0.3661 | 0.2316 | 0.2775 | 0.4342 | 0.4336 | 0.4338 | 0.4336 |
| 2010 | 0.2921 | 0.1700 | 0.2094 | 0.3173 | 0.3171 | 0.3173 | 0.3175 |
| 2011 | 0.1961 | 0.1090 | 0.1329 | 0.1894 | 0.1894 | 0.1894 | 0.1893 |
| 2012 | 0.1803 | 0.0918 | 0.1200 | 0.1658 | 0.1658 | 0.1658 | 0.1658 |
| 2013 | 0.1449 | 0.0706 | 0.0928 | 0.1205 | 0.1205 | 0.1205 | 0.1205 |
| 2014 | 0.1168 | 0.0556 | 0.0762 | 0.0906 | 0.0906 | 0.0905 | 0.0906 |
| 2015 | 0.1055 | 0.0527 | 0.0760 | 0.0836 | 0.0836 | 0.0836 | 0.0836 |
| 2016 | 0.0824 | 0.0407 | 0.0604 | 0.0613 | 0.0613 | 0.0613 | 0.0613 |
| 2017 | 0.0676 | 0.0329 | 0.0475 | 0.0471 | 0.0471 | 0.0471 | 0.0471 |
| 2018 | 0.0519 | 0.0262 | 0.0364 | 0.0347 | 0.0347 | 0.0347 | 0.0347 |
| 2019* | 0.0540 | 0.0272 | 0.0379 | 0.0362 | 0.0362 | 0.0362 | 0.0362 |
| 2020* | 0.0560 | 0.0283 | 0.0394 | 0.0376 | 0.0376 | 0.0376 | 0.0376 |
| GM(81-18) | 0.2081 | 0.2292 | 0.3071 | 0.4243 | 0.4242 | 0.4242 | 0.4243 |

* Estimates for 2019 and 2020 are TSA projections.

Table 38.15. Whiting in Division 6.a. Standard errors of TSA estimates for $\log$ mortality-at-age.

| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 0.0132 | 0.0141 | 0.0240 | 0.0333 | 0.0333 | 0.0333 | 0.0333 |
| 1982 | 0.0171 | 0.0190 | 0.0300 | 0.0365 | 0.0365 | 0.0365 | 0.0365 |
| 1983 | 0.0290 | 0.0332 | 0.0499 | 0.0536 | 0.0535 | 0.0535 | 0.0535 |
| 1984 | 0.0377 | 0.0473 | 0.0631 | 0.0636 | 0.0639 | 0.0635 | 0.0639 |
| 1985 | 0.0410 | 0.0526 | 0.0694 | 0.0715 | 0.0715 | 0.0715 | 0.0713 |
| 1986 | 0.0335 | 0.0443 | 0.0564 | 0.0571 | 0.0572 | 0.0571 | 0.0571 |
| 1987 | 0.0417 | 0.0527 | 0.0649 | 0.0663 | 0.0664 | 0.0664 | 0.0664 |
| 1988 | 0.0477 | 0.0643 | 0.0703 | 0.0801 | 0.0799 | 0.0798 | 0.0798 |
| 1989 | 0.0431 | 0.0606 | 0.0662 | 0.0729 | 0.0728 | 0.0729 | 0.0729 |
| 1990 | 0.0333 | 0.0441 | 0.0521 | 0.0569 | 0.0568 | 0.0568 | 0.0568 |
| 1991 | 0.0339 | 0.0475 | 0.0525 | 0.0588 | 0.0589 | 0.0590 | 0.0589 |
| 1992 | 0.0329 | 0.0469 | 0.0535 | 0.0610 | 0.0610 | 0.0611 | 0.0610 |
| 1993 | 0.0343 | 0.0482 | 0.0567 | 0.0741 | 0.0739 | 0.0742 | 0.0739 |
| 1994 | 0.0318 | 0.0438 | 0.0531 | 0.0702 | 0.0693 | 0.0696 | 0.0697 |
| 1995 | 0.0380 | 0.0518 | 0.0615 | 0.0847 | 0.0846 | 0.0847 | 0.0844 |
| 1996 | 0.0521 | 0.0682 | 0.0789 | 0.1067 | 0.1069 | 0.1067 | 0.1068 |
| 1997 | 0.0613 | 0.0804 | 0.0871 | 0.1122 | 0.1119 | 0.1124 | 0.1120 |
| 1998 | 0.0699 | 0.0889 | 0.0912 | 0.1091 | 0.1087 | 0.1092 | 0.1092 |
| 1999 | 0.0837 | 0.1017 | 0.1024 | 0.1197 | 0.1199 | 0.1197 | 0.1196 |
| 2000 | 0.0838 | 0.0937 | 0.0949 | 0.1240 | 0.1241 | 0.1240 | 0.1241 |
| 2001 | 0.0795 | 0.0845 | 0.0864 | 0.1195 | 0.1195 | 0.1192 | 0.1195 |
| 2002 | 0.0653 | 0.0679 | 0.0672 | 0.0933 | 0.0932 | 0.0933 | 0.0934 |
| 2003 | 0.0757 | 0.0741 | 0.0705 | 0.1005 | 0.1003 | 0.1006 | 0.1003 |
| 2004 | 0.0932 | 0.0831 | 0.0870 | 0.1156 | 0.1161 | 0.1154 | 0.1155 |
| 2005 | 0.0918 | 0.0753 | 0.0834 | 0.1113 | 0.1112 | 0.1106 | 0.1107 |
| 2006 | 0.0744 | 0.0503 | 0.0492 | 0.0637 | 0.0629 | 0.0629 | 0.0634 |
| 2007 | 0.0677 | 0.0409 | 0.0396 | 0.0575 | 0.0570 | 0.0583 | 0.0576 |


| Year | Age <br> 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 0.0791 | 0.0487 | 0.0431 | 0.0611 | 0.0611 | 0.0609 | 0.0606 |
| 2009 | 0.0768 | 0.0433 | 0.0407 | 0.0525 | 0.0533 | 0.0528 | 0.0532 |
| 2010 | 0.0619 | 0.0326 | 0.0313 | 0.0409 | 0.0408 | 0.0412 | 0.0416 |
| 2011 | 0.0429 | 0.0217 | 0.0210 | 0.0248 | 0.0248 | 0.0248 | 0.0248 |
| 2012 | 0.0413 | 0.0191 | 0.0198 | 0.0222 | 0.0222 | 0.0221 | 0.0222 |
| 2013 | 0.0347 | 0.0152 | 0.0162 | 0.0164 | 0.0165 | 0.0164 | 0.0165 |
| 2014 | 0.0284 | 0.0124 | 0.0138 | 0.0122 | 0.0122 | 0.0122 | 0.0122 |
| 2015 | 0.0266 | 0.0123 | 0.0145 | 0.0114 | 0.0114 | 0.0114 | 0.0114 |
| 2016 | 0.0218 | 0.0100 | 0.0126 | 0.0086 | 0.0086 | 0.0086 | 0.0086 |
| 2017 | 0.0190 | 0.0087 | 0.0107 | 0.0068 | 0.0068 | 0.0068 | 0.0068 |
| 2018 | 0.0157 | 0.0075 | 0.0091 | 0.0053 | 0.0053 | 0.0053 | 0.0053 |
| 2019* | 0.0219 | 0.0108 | 0.0141 | 0.0116 | 0.0116 | 0.0116 | 0.0116 |
| 2020* | 0.0282 | 0.0140 | 0.0189 | 0.0168 | 0.0168 | 0.0168 | 0.0168 |
| GM(81-18) | 0.0432 | 0.0388 | 0.0438 | 0.0498 | 0.0498 | 0.0498 | 0.0498 |

[^23]Table 38.16. Whiting in Division 6a. TSA summary table. "Obs." denotes sum-of-products of numbers and mean weights-at-age, not reported caught, landed and discarded weight.

| Year | Landings (tonnes) |  |  | Discards (tonnes) |  |  | Total catches (tonnes)) |  |  | Mean F(2-4) |  | SSB (tonnes) |  | TSB (tonnes) |  | Recruitment (000s at age 1) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Obs. | Pred. | SE | Obs. | Pred. | SE | Obs. | Pred. | SE | Estimate | SE | Estimate | SE | Estimate | SE | Estimate | SE |
| 1981 | 12194 | 11441 | 1281 | 2132 | 4554 | 948 | 14325 | 15994 | 1465 | 0.224 | 0.020 | 134180 | 7515 | 168467 | 8465 | 199423 | 20088 |
| 1982 | 13880 | 13052 | 1425 | 5485 | 4362 | 921 | 19366 | 17415 | 1637 | 0.253 | 0.023 | 91841 | 4952 | 109621 | 5444 | 165088 | 17707 |
| 1983 | 15962 | 16708 | 1550 | 6294 | 5285 | 935 | 22257 | 21994 | 2014 | 0.418 | 0.035 | 63123 | 3555 | 92947 | 4908 | 197237 | 19493 |
| 1984 | 16459 | 14390 | 1305 | 4017 | 5084 | 951 | 20476 | 19474 | 1870 | 0.537 | 0.044 | 46147 | 2889 | 81788 | 4428 | 325817 | 26806 |
| 1985 | 12879 | 11380 | 1113 | 4840 | 7183 | 1245 | 17719 | 18563 | 1810 | 0.619 | 0.048 | 42257 | 2699 | 78815 | 4152 | 310357 | 24040 |
| 1986 | 8458 | 7875 | 847 | 2669 | 5362 | 915 | 11127 | 13237 | 1325 | 0.484 | 0.040 | 38271 | 2467 | 72299 | 3751 | 291281 | 21630 |
| 1987 | 11542 | 9953 | 996 | 11918 | 8082 | 1377 | 23460 | 18035 | 1789 | 0.583 | 0.045 | 40750 | 2422 | 77744 | 4175 | 404905 | 34057 |
| 1988 | 11349 | 10568 | 1013 | 8132 | 5496 | 1068 | 19481 | 16063 | 1522 | 0.678 | 0.052 | 41444 | 2550 | 50179 | 2945 | 106310 | 11437 |
| 1989 | 7523 | 6629 | 705 | 5876 | 6122 | 1058 | 13399 | 12750 | 1412 | 0.603 | 0.049 | 22737 | 1695 | 57007 | 3091 | 328200 | 22212 |
| 1990 | 5642 | 5235 | 561 | 4530 | 4944 | 910 | 10172 | 10180 | 1123 | 0.436 | 0.039 | 33795 | 2077 | 57751 | 3593 | 174358 | 18899 |
| 1991 | 6658 | 5750 | 558 | 4883 | 4022 | 740 | 11541 | 9773 | 1029 | 0.448 | 0.040 | 27593 | 1873 | 51889 | 3282 | 244996 | 23526 |
| 1992 | 6005 | 5696 | 539 | 9249 | 6124 | 1064 | 15253 | 11820 | 1292 | 0.437 | 0.042 | 30711 | 2162 | 68550 | 4423 | 334144 | 29277 |
| 1993 | 6872 | 6735 | 645 | 4759 | 7036 | 1175 | 11631 | 13771 | 1418 | 0.463 | 0.048 | 43157 | 3099 | 75332 | 5199 | 262160 | 24459 |
| 1994 | 5901 | 5951 | 572 | 3455 | 5177 | 812 | 9356 | 11128 | 1070 | 0.416 | 0.045 | 37684 | 3110 | 62929 | 4984 | 274566 | 27139 |
| 1995 | 6078 | 6758 | 1060 | 5771 | 5782 | 1040 | 11849 | 12540 | 1836 | 0.463 | 0.055 | 37485 | 3655 | 59706 | 5076 | 291943 | 25465 |
| 1996 | 7158 | 7611 | 1326 | 7940 | 7364 | 1421 | 15098 | 14975 | 2442 | 0.572 | 0.072 | 40626 | 3808 | 58196 | 5308 | 179537 | 20640 |


| Year | Landings (tonnes) |  |  | Discards (tonnes) |  |  | Total catches (tonnes)) |  |  | Mean F(2-4) |  | SSB (tonnes) |  | TSB (tonnes) |  | Recruitment (000s at age 1) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Obs. | Pred. | SE | Obs. | Pred. | SE | Obs. | Pred. | SE | Estimate | SE | Estimate | SE | Estimate | SE | Estimate | SE |
| 1997 | 6290 | 7635 | 1197 | 5251 | 6599 | 1319 | 11542 | 14234 | 2230 | 0.632 | 0.080 | 31884 | 3215 | 50755 | 5358 | 163752 | 23369 |
| 1998 | 4627 | 5531 | 919 | 9216 | 7289 | 1568 | 13843 | 12820 | 2235 | 0.705 | 0.081 | 22259 | 2994 | 43695 | 5808 | 214790 | 33450 |
| 1999 | 4613 | 4755 | 942 | 3975 | 6261 | 1466 | 8588 | 11016 | 2191 | 0.848 | 0.088 | 20081 | 3341 | 33648 | 5333 | 152810 | 27455 |
| 2000 | 3011 | 3414 | 790 | 13285 | 6481 | 1577 | 16296 | 9895 | 2176 | 0.814 | 0.085 | 15043 | 2883 | 32807 | 5351 | 235620 | 38631 |
| 2001 | 2439 | 2971 | 648 | 4263 | 5096 | 1214 | 6702 | 8067 | 1713 | 0.723 | 0.079 | 16962 | 2919 | 27227 | 4223 | 102705 | 15914 |
| 2002 | 1767 | 2351 | 551 | 2851 | 1960 | 530 | 4618 | 4311 | 998 | 0.529 | 0.063 | 12473 | 2082 | 15480 | 2578 | 40754 | 8708 |
| 2003 | 1355 | 1902 | 439 | 719 | 1776 | 501 | 2074 | 3679 | 869 | 0.562 | 0.068 | 7916 | 1363 | 13248 | 2154 | 63845 | 11261 |
| 2004 | 811 | 1125 | 275 | 2159 | 1649 | 508 | 2970 | 2774 | 738 | 0.585 | 0.082 | 5791 | 1050 | 9266 | 1551 | 40215 | 7129 |
| 2005 | 341 | 701 | 174 | 629 | 875 | 255 | 970 | 1576 | 402 | 0.487 | 0.080 | 3814 | 569 | 5903 | 771 | 23483 | 3420 |
| 2006 | 380 | 542 | 54 | 946 | 644 | 114 | 1327 | 1186 | 143 | 0.410 | 0.041 | 3481 | 243 | 4819 | 307 | 28198 | 2578 |
| 2007 | 427 | 441 | 39 | 317 | 451 | 81 | 745 | 892 | 101 | 0.325 | 0.035 | 3445 | 210 | 4737 | 297 | 15449 | 1852 |
| 2008 | 445 | 436 | 40 | 314 | 554 | 100 | 759 | 990 | 120 | 0.366 | 0.039 | 3093 | 238 | 4385 | 311 | 16893 | 1685 |
| 2009 | 488 | 419 | 40 | 419 | 516 | 91 | 908 | 935 | 113 | 0.314 | 0.035 | 3497 | 306 | 4836 | 369 | 25105 | 2107 |
| 2010 | 307 | 309 | 32 | 893 | 565 | 103 | 1200 | 874 | 117 | 0.232 | 0.027 | 2809 | 249 | 5121 | 384 | 59545 | 5375 |
| 2011 | 230 | 252 | 27 | 339 | 322 | 58 | 569 | 573 | 71 | 0.144 | 0.018 | 5673 | 484 | 6245 | 509 | 19074 | 1703 |
| 2012 | 313 | 279 | 31 | 727 | 473 | 86 | 1039 | 752 | 98 | 0.126 | 0.016 | 5499 | 479 | 7700 | 638 | 38435 | 4454 |
| 2013 | 222 | 237 | 25 | 951 | 289 | 51 | 1173 | 526 | 63 | 0.095 | 0.013 | 6497 | 624 | 7313 | 677 | 19782 | 2168 |


| Year | Landings (tonnes) |  |  | Discards (tonnes) |  |  | Total catches (tonnes)) |  |  | Mean F(2-4) |  | SSB (tonnes) |  | TSB (tonnes) |  | Recruitment (000s at age 1) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Obs. | Pred. | SE | Obs. | Pred. | SE | Obs. | Pred. | SE | Estimate | SE | Estimate | SE | Estimate | SE | Estimate | SE |
| 2014 | 184 | 210 | 21 | 583 | 300 | 60 | 767 | 511 | 70 | 0.074 | 0.010 | 6378 | 632 | 8612 | 846 | 45389 | 6285 |
| 2015 | 227 | 210 | 21 | 835 | 702 | 147 | 1063 | 912 | 155 | 0.071 | 0.010 | 7328 | 838 | 15178 | 1422 | 105619 | 10118 |
| 2016 | 233 | 232 | 24 | 797 | 591 | 115 | 1030 | 823 | 127 | 0.054 | 0.008 | 15680 | 1571 | 20569 | 2065 | 66931 | 9868 |
| 2017 | 176 | 209 | 23 | 1207 | 565 | 116 | 1383 | 774 | 125 | 0.043 | 0.007 | 16824 | 1813 | 22031 | 2738 | 79508 | 19510 |
| 2018 | 189 | 206 | 23 | 492 | 342 | 86 | 681 | 548 | 96 | 0.032 | 0.006 | 17175 | 2547 | 21136 | 3549 | 60000 | 23195 |
| 2019* | NA | 275 | 83 | NA | 553 | 219 | NA | 829 | 283 | 0.034 | 0.011 | 18961 | 3643 | 28516 | 5807 | 140154 | 47647 |
| 2020* | NA | 327 | 136 | NA | 759 | 376 | NA | 1085 | 491 | 0.035 | 0.015 | 25931 | 5960 | 37869 | 8812 | 175106 | 69044 |
| Min | 176 | 206 | 21 | 314 | 289 | 51 | 569 | 511 | 63 | 0.032 | 0.006 | 2809 | 210 | 4385 | 297 | 15449 | 1685 |
| GM | 1929 | 2038 | 252 | 2237 | 2122 | 423 | 4547 | 4310 | 585 | 0.321 | 0.035 | 16528 | 1514 | 25878 | 2206 | 100335 | 11783 |
| AM | 4833 | 4739 | 575 | 3766 | 3586 | 703 | 8599 | 8326 | 1053 | 0.416 | 0.044 | 26458 | 2136 | 41788 | 3170 | 150216 | 16503 |
| Max | 16459 | 16708 | 1550 | 13285 | 8082 | 1577 | 23460 | 21994 | 2442 | 0.848 | 0.088 | 134180 | 7515 | 168467 | 8465 | 404905 | 38631 |

[^24]

Figure 38.1. Whiting in Division 6.a. Landings, discards and catch (in tonnes, whiting at-age 1 and older) as officially reported to ICES (upper panel) and discards (as \% of catch, lower panel).

whg.27.6a DisWt


Figure 38.2. Whiting in Division 6.a. Landings (upper panel) and discards (all ages, lower panel) by métier (kg) in 2019 as entered into InterCatch.

Landing weights at age for whiting in 27.6.a


Discard weights at age for whiting in 27.6.a


Catch weights at age for whiting in 27.6.a


Figure 38.3. Whiting in Division 6.a. Mean weight-at-age in the landings (upper panel), discards (middle panel) and catch (lower panel).

UK-SCOWCGFS-Q1 \& UK-SCOWCGFS-Q4: whiting


Figure 38.4. Whiting in Division 6.a. Cpue from the Scottish first quarter west coast groundfish survey (UK-SCOWCGFS-Q1, in red) and the Scottish fourth quarter groundfish survey (UK-SCOWCGFS-Q4, in blue) in 2016-2019. Numbers are standardised to 30 minutes towing. Two closed areas (the Windsock in the north and the Clyde in the south) are shown as green polygons.

Figure 38.5. Whiting in Division 6.a. Cpue from the Scottish fourth quarter west coast groundfish survey (UK-SCOWCGFS-Q4, only the southern part of the survey area, in blue) and the Irish fourth quarter groundfish survey (IGFS-WIBTS-Q4, in green) in 2015-2018. Numbers are standardised to 30 minutes towing. The Clyde closed area is shown as a green polygon.


Figure 38.6. Whiting in Division 6.a. Scaled survey indices (Z-scores) from ScoGFS-WIBTS-Q1, ScoGFS-WIBTS-Q4, IGFS-WIBTS-Q4, UK-SCOWCGFS-Q1 and UK-SCOWCGFS-Q4. The abundance index for IGFS-WIBTS-Q4 is shown only for ages 0-6.


Figure 38.7. Whiting in Division 6.a. Log mean standardised survey index for each age by cohort (upper panels) and year (lower panels) in IGFS-WIBTS-Q4, UK-SCOWCGFSQ1 and UK-SCOWCGFS-Q4, respectively.


Figure 38.8. Whiting in Division 6.a. Log catch curves from the catch (ages 1-7+) and from the five survey series (ages as specified in Table 38.9).


Figure 38.9. Whiting in Division 6.a. Proportion discarded at-age from the final TSA run.


Figure 38.10. Whiting in Division 6.a. Standardised residuals for landings (left panel) and discards (right panel) from the final TSA run.


Figure 38.11. Whiting in Division 6.a. Standardised survey residuals from TSA in ScoGFS-WIBTS-Q1 (top left panel), ScoGFS-WIBTS-Q4 (top left panel), IGFS-WIBTS-Q4 (middle panel), UK-SCOWCGFS-Q1 (bottom left panel) and UK-SCOWCGFS-Q4 (bottom right panel), from the final TSA run.


Figure 38.12. Whiting in Division 6.a. Percentage change in catchability from the final TSA run. Transient changes (points) and the persistent change (solid line) with uncertainty bounds.


Figure 38.13. Whiting in Division 6.a. Stock-recruitment relationship (recruitment in millions, SSB in thousand tonnes) from the final TSA run, with points labelled as year classes, and fitted with a segmentedregression model ("hockey-stick", solid line).


Figure 38.14. Whiting in Division 6.a. TSA stock summaries from the final TSA run. Catch, landings, discards and SSB in tonnes, recruitment in thousands. Estimates are plotted with approximate pointwise $95 \%$ confidence bounds. Dots indicate observed values for catch, landings and discards.


Figure 38.15. Whiting in Division 6.a. Retrospective plots of TSA run (the retro analysis for 2009-2018). Catch, landings, discards and SSB in tonnes, recruitment in thousands. Blue points show observed values, black lines show estimates in the respective years, grey bands show confidence intervals for the last estimate.

## 37 Whiting (Merlangius merlangus) in Division 6.b (Rockall)

## Type of assessment in 2019

No assessment was performed in 2018.

## ICES advice applicable to 2019

In 2018, ICES provided multiyear advice:
ICES advises that when the precautionary approach is applied, wanted catches should be no more than 9 tonnes in each of the years 2019, 2020, and 2021. ICES cannot quantify the corresponding total catches.
http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2018/2018/whg.27.6b.pdf

## ICES advice applicable to 2016-2018

In 2015, ICES provided multiyear advice:
ICES advises that when the precautionary approach is applied, catches should be no more than 11 tonnes in each of the years 2016, 2017, and 2018.
http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2015/2015/whg-rock.pdf

### 37.1 General

## Stock description

There is an absence of information on whiting stock structure in this region and whiting caught at Rockall may potentially be part of the adjacent 6 .a stock.

## Management applicable to 2018 and 2019

The TAC for whiting (in tonnes) is set for ICES subareas 6,12 , and 14 , and EU and international waters of ICES Division 5b, for 2019 and 2018 is shown below.

TAC for 2019

| Species: | Whiting <br> Merlangius merlangus |  | Zone: | 6; Union and international waters of 5 b ; international waters of 12 and 14 <br> (WHG/56-14) |
| :---: | :---: | :---: | :---: | :---: |
| Germany |  | $3{ }^{(1)}$ |  |  |
| France |  | $68\left({ }^{1}\right)$ |  |  |
| Ireland |  | $324{ }^{(1)}$ |  |  |
| United Kingdom |  | $717{ }^{1}{ }^{1}$ |  |  |
| Union |  | $1112{ }^{(1)}$ |  |  |
| TAC |  | $1112{ }^{(1)}$ |  | Analytical TAC <br> Article 8 of this Regulation applies |

[^25]TAC for 2018

| Species: | Whiting Merlangius merlangus |  | Zone: | 6; Union and international waters of 5 b ; international waters of 12 and 14 (WHG/56-14) |
| :---: | :---: | :---: | :---: | :---: |
| Germany |  | $1{ }^{1}{ }^{1}$ |  |  |
| France |  | $26{ }^{(1)}$ |  |  |
| Ireland |  | $64\left({ }^{1}\right)$ |  |  |
| United Kingdom |  | $122{ }^{(1)}$ |  |  |
| Union |  | $213{ }^{(1)}$ |  |  |
| TAC |  | $213{ }^{(1)}$ |  | Analytical TAC |
| ${ }^{\left({ }^{1}\right)}$ Exclusively for by-catches. No directed fisheries are permitted under this quota. |  |  |  |  |

(Council Regulation (EU) 2018/120).

## Fishery in 2018

No specific information is available for 2018. Whiting at Rockall are taken as a bycatch in fisheries for other species such as haddock and anglerfish.

### 37.2 Data

Landings data for whiting in $6 . b$ are shown by nation in Table 39.1 and Figure 39.1. Total officially reported landings were 43 t in 2018, of which 34 t were reported by the UK, 9 t by Ireland, and $<0.5 \mathrm{t}$ by Norway. In the past, official landings have shown very high interannual variation and it is not known whether these are a true reflection of removals.

Landings and discards have been uploaded to InterCatch for 2018 (Figure 39.2).
In addition, some landings and discards age compositions were also uploaded to InterCatch. About $78 \%$ of the total landings ( 44 t ) are from the Scottish TR1 fleet which, based on two sampled trips has a $12 \%$ discard rate. The data available in InterCatch are shown below.

| Country | Landings ( $\mathbf{t}$ ) | Discards ( $\mathbf{t}$ ) | Total ( $\mathbf{t}$ ) |
| :--- | :---: | :---: | :---: |
| Ireland | 9.0 | 1.2 | 10.2 |
| UK (Scotland) | 34.0 | 4.6 | 38.6 |
| Norway | 0.5 | 0.1 | 0.6 |
| Grand total | 5.9 | 43.6 | 49.4 |

Survey catch rates of whiting at Rockall are extremely low (Table 39.2, Figure 39.3) and are therefore unlikely to provide a reliable index of abundance.
Catches of whiting (both survey and commercial) are too low to support the collection of the necessary information for an assessment of stock status.

### 37.3 Target category

In 2012, advice was provided using the DL approach for category 6.2.0; stocks with negligible landings stocks and stocks caught in minor amounts as bycatch with no indication of F in relation
to reference points and no marked positive trends in stock indicators. WKLIFE has previously suggested a target category of 4 for this stock. Given the comments in Section 39.2 regarding the potential unreliability of landings data and lack of sampled data, WGCSE considers that whiting in $6 . b$ is likely to remain a category 6 stock.

### 37.4 Management considerations

Rockall whiting is managed under a TAC for the combined Divisions 6.a and 6.b, and therefore cannot be effective in limiting catches in Rockall.

Table 39.1. Whiting in Division 6.b. Nominal landings (in tonnes) as officially reported to ICES.

| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| France | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Ireland | - | - | - | - | 32 | 10 | 4 | 23 | 3 | 1 | - | - | 10 | - | 2 | 3 | 3 | 104 |
| Norway | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Spain | - | - | - | - | - | - | - | - | - | - | + | - | - | - | - | - | - | - |
| UK (E, W \& NI) | 16 | 6 | 1 | 5 | 10 | 2 | 5 | 26 | 49 | 20 | - | - | - | - | - | - | - | - |
| UK (Scotland) | 18 | 482 | 459 | 283 | 86 | 68 | 53 | 36 | 65 | 23 | 44 | 58 | 4 | 7 | 11 | 1 | 1 | 1 |
| UK (all) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 34 | 488 | 460 | 288 | 128 | 80 | 62 | 85 | 117 | 44 | 44 | 58 | 14 | 7 | 13 | 4 | 4 | 105 |
| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017* | 2018* |  |  |  |  |  |  |
| Faroe Islands | - | - | - | - | - | - | - | - | - | - | - | - |  |  |  |  |  |  |
| France | + | - | - | - | - | - | - | - | - | - | - | - |  |  |  |  |  |  |
| Ireland | 16 | 23 | 4 | 2 | 3 | - | + | 6 | 6 | 9 | 7 | 9 |  |  |  |  |  |  |
| Norway | - | - | - | - | - | - | - | - | - | 1 | - | + |  |  |  |  |  |  |
| Spain | - | - | - | - | - | - | - | - | - | - | - | - |  |  |  |  |  |  |
| UK (E, W \& NI) | - | - | - | - | - | - | - | - |  |  |  |  |  |  |  |  |  |  |
| UK (Scotland) | 1 | 8 | 12 | 16 | 6 | 1 | 3 | 23 |  |  |  |  |  |  |  |  |  |  |
| UK (all) |  |  |  |  |  |  |  |  | 46 | 22 | 32 | 34 |  |  |  |  |  |  |
| Total | 17 | 31 | 16 | 18 | 9 | 1 | 3 | 29 | 52 | 33 | 40 | 43 |  |  |  |  |  |  |

* Preliminary.
$+<0.5 \mathrm{t}$.

Table 39.2. Whiting in Division 6.b. Survey data made available to the WG: Scottish Q3 groundfish survey (UK-SCOROC-Q3). Catch rates are given as number per ten hours.

| UK-SCOROC-Q3 - SCOTTISH GROUNDFISH SURVEY - NUMBERS AT AGE/10 H |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Effort | Age |  |  |  |  |  |  |  |
|  | (hours) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 2011 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 10 | 33.279 | 0 | 0.358 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 10 | 6.687 | 1.924 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 10 | 17.425 | 3.426 | 0.838 | 0.307 | 0 | 0 | 0 | 0 |
| 2015 | 10 | 8.853 | 0.559 | 0.559 | 0.55 | 0 | 0 | 0 | 0 |
| 2016 | 10 | 250.012 | 0.782 | 0 | 0.223 | 0.447 | 0 | 0 | 0 |
| 2017 | 10 | 23.147 | 10.84 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 10 | 0.531 | 0.754 | 0.894 | 0 | 0 | 0.307 | 0 | 0 |



Figure 39.1. Whiting in Division 6.b. Official landings by nation.
whg.27.6b LandWt

whg.27.6b DisWt


Figure 39.2. Whiting in Division 6.b. Landings (upper panel) and discards (lower panel), all ages, by métier $(\mathrm{kg})$ in 2018 as entered into InterCatch.


Figure 39.3. Whiting in Division 6.b. Whiting distribution by age on the Rockall Bank in 2012-2018 as observed in the Rockall Haddock survey. The densities (numbers of fish per 30 min ) are represented by circles. The red polygons show the protected areas. The red rectangle in the centre shows the Haddock Box. The dashed line shows the NEAFC Regulatory Area.

## 38 Whiting (Merlangius merlangus) in Division 7.a (Irish Sea)

## 2019 Assessment and advice

This stock was benchmarked in 2017 and the outcome was to upgrade the assessment from category 3 (trends based) to category 1 (analytical assessment and forecast). Data exploration was carried out in WKIrish2 (ICES, 2017). A full analytical assessment procedure was developed during WKIrish3 (ICES, 2017) using ASAP. Reference points were also estimated during WKIrish3. WGCSE 2019 updated the assessment with 2018 data. The advice for this stock is biennial and is produced for 2019; however, it was also updated in October 2018 following a special request to ICES to update the advice based on the most recent discard estimates.

Furthermore, in response to an EC request for advice on the removal of TACs for certain stocks ICES advises that removing the EU TAC for Whiting in ICES Division 7.a may generate a high risk of the stock being unsustainably exploited. However, ICES notes that the TAC is not currently controlling exploitation.

## Type of assessment

SPALY update of ASAP assessment.

## ICES advice applicable to 2019 and 2020

ICES advises that when the MSY approach is applied, there should zero catches in each of the years 2018 and 2019.
http://ices.dk/sites/pub/Publication\ Reports/Advice/2018/Special requests/uk.2018.29.pdf

### 38.1 General

## Stock description and management units

The stock and the management unit are both ICES Division 7.a (Irish Sea). Whiting landings taken or reported in ICES rectangles 33E2 and 33E3 have been reassigned to the 7.b,c,e-k whiting stock since 2003.


## Management applicable to 2017 and 2018

The minimum conservation reference size of whiting is 27 cm . The 2018 TAC for whiting 7.a was 80 t . Overall, official landings in 2018 were below the TAC but some countries landings were
close to their quotas. This stock is now subject to the landings obligation by way of the Commission Delegated Regulation (EU) 2018/2034. In 2019, the TAC was increased to 727 t .

| 2018 | 2018 Quota | 2018 Officially reported Landings |
| :--- | :---: | :---: |
| Belgium | 0 | 1 |
| France | 3 | $<0.5$ |
| Ireland | 46 | 44 |
| The Netherlands | 0 | - |
| United Kingdom | 31 | 19 |

Note for Ireland, 18 t were reallocated from rectangles $33 \mathrm{E} 2 \& 33 \mathrm{E} 3$.

TAC 2018

| Species:Whiting <br> Merlangius merlangus | Zone:VIla <br> (WHG/07A.) |  |
| :--- | :---: | :---: | :---: |
| Belgium | 0 |  |
| France | 46 |  |
| Ireland | 0 |  |
| The Netherlands | 31 | Analytical TAC |
| United Kingdom | 80 |  |
| Union | 80 |  |
| TAC |  |  |

## TAC 2019



## Fishery in 2018

The characteristics of the fishery are described in the stock annex.
The fishery in 2018 was prosecuted by the same fleets and gears as in recent years. In addition to this, $65 \%$ of the Irish landings were submitted were from the PTM_SPF métier for 2018. These were from trips targeting herring where whiting was a bycatch.

Table 40.1 gives the official nominal landings of 7 .a whiting as reported by each country to ICES. Working Group estimates of the landings and discards for the main fleets are given in Table 40.2. In recent years, the values provided to the WG are very similar to officially reported landings.

The majority of the catch was discarded in the Nephrops fishery (853 t) by UK-NI and IRE (Table 40.2).

The closure of the western Irish Sea to whitefish fishing from mid-February to the end of April, designed to protect cod, was continued in 2018 but is unlikely to have affected whiting catches, which are mainly bycatch in the derogated Nephrops fishery. Nephrops vessels can obtain a derogation to fish in certain sections of the closed area, providing they fit separator panels to their nets to allow escape of cod and other fish. The TR2 fleet in 7.a are obliged to use one of four types of cod selective measures, namely a 'Swedish' grid; the inclined separator panel, SELTRA trawl or 300 square mesh panel.

A summary of the 2018 catches by main gear types is presented below.

| Catch (2018) | Landings |  | Discards |  |  |
| :---: | :--- | :--- | :--- | :--- | :--- |
|  | fin-fish trawls | Nephrops directed otter <br> trawl | Other gears | Nephrops directed otter <br> trawls | Other gears |
|  |  | $14.3 \%$ | $48 \%$ | $98 \%$ | $2 \%$ |
|  | $38 \%$ | 46 t | 853 t |  |  |

### 38.2 Information from the Industry

There was no information on the whiting stock from the industry.

### 38.3 Data

Data were provided by all countries according to the data call.
For WGCSE (2019) all data have been updated. To allow an age-based assessment, catch num-bers-at-age, catch weights-at-age, stock weights-at-age have all be constructed since 2003 (WGCSE, 2017). These updates are documented in the Stock Annex.

## Fishery landings

Working Group estimates of catch available since 1980 are illustrated in Figure 40.1 and indicate the declining trend since the start of the time-series.

The introduction of UK and Irish legislation requiring registration of fish buyers and sellers may mean that the reported landings from 2006 onwards are more representative of actual landings.
Working group estimates of landings are corrected for misreporting in the past. There is information that officially reported landings of whiting, especially around the mid-1990s, have been inaccurate due to misreporting. Landings data have previously been partially corrected for by using sample-based estimates of landings at a number of Irish Sea ports. Due to the low level of landings recently, this has not been carried out since 2003. As for 7.a cod and haddock, the whiting landings taken or reported in ICES rectangles 33 E 2 and 33E3 have been reassigned to the 7.ek whiting stock since 2003 (Table 40.3).

## Fishery discards

Discard estimates are available from Ireland and Northern Ireland, with minor discards from the UK E\&W. Data were submitted for the OTB_DEF_100-119_0_0_all métier from Ireland due to increased sampling in this métier in 2018. Raising methods used are described in the stock annex for 7.a whiting.

## Landings-at-age data

Sampling and raising methods previously used are described in the stock annex for 7.a whiting. Methods for estimating quantities and composition of landings are described in the stock annex.

Landings numbers-at-age are given in Table 40.4. For the 2003 data onwards, the catch and mean weight-at-age are estimated using combined UK (NI) and Irish quarterly length-weight relationships and age-length keys. These data are raised to the international catch data provided to ICES. Typically, quarterly landings are provided by the UK (Scotland), Belgium and France and annual landings are provided by UK (IOM). The quality of the landings-at-age data has been declining in recent years due to reduced sample numbers commensurate with the decline in landings.

## Discards numbers-at-age data

Discard number-at-age are given in Table 40.5. Discarding of whiting is high within the Irish Sea. Discard Numbers-at-age were combined for Ireland and Northern Ireland for ages 0 to $6+$ and then raised to the international discards. From 2003, the discard time-series from Ireland is based on the Nephrops fleet only. Therefore, the discard weight in tonnes has been revised. Discards from NI were not available from 2003-2005 and so discard numbers-at-age are based on Irish sampling data only. There has been a high number of age 1 and 2 discarded at the start of the time-series with almost all age 1 and 2 discarded later in time-series (Figure 40.3).

The length-frequency of discards of national sampled fleets in 2018 is given in Figure 40.2. More detailed information is available in the stock annex.

## Biological data

The derivation of these parameters and variables is described in the stock annex. The Lorenzen method was used to estimate M . This was derived during WKIrish2 and investigated during WKIrish3. Maturity-at-age is knife-edge at-age 2. Stock weights were also revised at the benchmark meeting. Stock weights-at-age were derived from the catch weights and then smoothed using a three-year moving average. Figure 40.4 shows the stock weights used. There are strong trends in mean weights-at-age over the time-series with a minimum around 2000s for most ages. There was a small increase in the mid-2000s but overall mean weights are significantly lower than at the start of the series.

## Survey data used in assessment

Table 40.6 describes the survey data made available to the Working Group.
In 2016, the entire time-series of the UK (E\&W)-BTS-Q3 survey data was revised so that only the selected prime stations are used.

Survey series for whiting provided to the Working Group are further described in the stock annex for 7.a whiting (Section B.3). Five survey series were available. The inclusion of the different available surveys was tested in a series of preliminary model runs at WKIrish3. Figure 40.5 shows the log-standardized indices of tuning fleets by cohort. There are very little cohort signals in any of the indices. The beam trawl survey shows an increasing trend in the early part of the time-series not seen in the other surveys. The three surveys included in the final assessment were NIGFSQ1, NIGFSQ2 and the NIMIK net survey.

### 38.4 Historical stock development

Model used: ASAP
Software used: ASAP V3.0.17 NOAA Fisheries toolbox (http://nft.nefsc.noaa.gov)

FLR with R version 3.1.2 with packages FLCore 2.5.20150309, FLAssess _2.5.20130716, FLXSA 2.5.20140808 and FLEDA 2.5 (http://flr-project.org)

## Data screening

The general approach to data screening and analysis was followed in addition to the data exploration tools available in the FLR package FLEDA. The results of the data screening are fully documented using R markdown and are available in the folder 'Data $\backslash$ Whg 7.a \Assessment. on SharePoint. Table 40.7 shows the ASAP input data.

## Final update assessment

The final assessment was run using the same settings as described in WKIrish3. These final settings are described in the Stock Annex.

The observed and predicted catches are shown in Figure 40.7. Fit to the overall catch is reasonably good. There is some deviation in the early to mid-1990s. This is most likely due to the introduction of the survey data into the assessment model.

The observed and predicted index cpue values are shown in Figure 40.8. There is poor fit to the Northern Irish groundfish survey indices in the first half of the series but it improves in recent years.

Figure 40.6 shows the selectivity-at-age in the catch. Full selectivity is assumed for age 3 and the model is allowed to estimate ages 1 and 2 . Table 40.8 shows the model estimates.

Figure 40.9 shows the retrospective analysis. The predicted catch shows no obvious retrospective pattern, neither does the recruitment estimate. There is some deviation in the early part of the time-series when the surveys were first introduced. However, recent estimates of SSB and F are consistent with no apparent bias.

A Mohn's rho analysis was conducted based on the ASAP stock assessment results, i.e. the last data year (2018) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (ages 1-3) | recruitment |
| :--- | :---: | :---: | :---: |
| Mohn's rho value | 0.129 | -0.190 | 0.29 |

The Mohn's rho values for this assessment are below the threshold of $20 \%$ imposed by ICES for 2019 assessments.

## The state of the stock

Table 40.9 shows the estimated fishing mortality-at-age and Table 40.10 shows the stock num-bers-at-age. The stock summary is given in Table 40.11 and Figure 40.10.

The present stock size is extremely low. SSB has declined since the start of the time-series and has been well below Blim since the mid-1990s. Recruitment has been low since the early 1990s. Large variations in fishing mortality estimates have been observed in recent years. F has been well above Flim since the early 1990s.

### 38.5 Short-term predictions

Short-term projections were performed using FLR libraries. Recruitment for 2019-2021was estimated at 118853 (GM 2000 onwards; thousands). Three-year averages were used for F (unscaled) and weights-at-age.

Input data for the short-term forecast are given in Table 40.12. The single-option output is given in Tables 40.13 and 40.14 gives the management options.

Estimates of the relative contribution of recent year classes to the 2020 landings and 2021 SSB are shown in Figure 40.11. The 2015-2017 year class estimates from ASAP accounts for 79\% of the projected landings in 2020. The 2019 GM assumption contributes considerably to the estimated SSB in 2021 as does the 2018 ASAP assessment.

### 38.6 Medium-term projection

There is no analytical assessment for this stock.

### 38.7 MSY evaluations and Biological Reference Points

ICES carried out and evaluation of MSY and PA reference points for this stock at WKIrish3. The results are summarized below:

|  | Type | Value | Technical basis |
| :--- | :--- | :--- | :--- |
| MSY | MSY Btrigger | 16300 t | Bpa |
| Approach | F MSY | FMSY lower | 0.22 |
| FMSY upper | 0.158 | Median point estimates of EqSim with combined SR |  |
| Precautionary | Bpa | 10000 t | Below 10,000 t recruitment is impaired |
| Approach | Flim | 16300 t | Blim combined with the assessment error |
|  | Fpa | 0.37 | F with 50\% probability of SSB less than Blim |

### 38.8 Management plans

No management plan has been agreed or proposed.

### 38.9 Uncertainties and bias in assessment and forecast

This stock was benchmarked in January 2017. The result of the benchmark was that the stock was elevated from a category 3 stock (trend-based assessment) to a category 1 stock (analytical assessment). The assessment includes information from the commercial fishery, including both landings and discards, and takes into account selectivity changes that have occurred in 1995. Three survey series are used within the assessment. Natural mortality parameters were updated to reflect current stock dynamics. The highly fluctuating estimates of fishing mortality in recent
years (2002-present) are likely to be the result of variability in the sampling data and discard estimates. Despite this inherent uncertainty, it is clear from the assessment and additional information from surveys that the stock remains extremely low.

Stock status classification relative to MSY proxies is given below.

|  | Fishing pressure |  |  |  |  | Stock size |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 2016 | 2017 |  | 2018 |  | 2017 | 2018 |  | 2019 |
| Maximum sustainable yield | $\mathrm{F}_{\text {MSY }}$ | ( | ( | ( | Above | MSY <br> $B_{\text {trigger }}$ | $x$ |  |  | Below trigger |
| Precautionary approach | $\mathrm{F}_{\mathrm{pa}} \mathrm{F}_{\mathrm{lim}}$ | ( | $\bigcirc$ | $x$ | Harvested unsustainably | $\mathrm{B}_{\mathrm{pa}} \mathrm{B} \mathrm{l}_{\text {lim }}$ | * | ( |  | Reduced reproductive capacity |
| Management plan | $\mathrm{F}_{\text {MGT }}$ | - | - | - | Not applicable | $\mathrm{B}_{\text {MGT }}$ | - | - |  | Not applicable |

### 38.10 Recommendations for next benchmark assessment

This stock was benchmarked in 2017 as part of the WKIRISH process. A number of recommendations for future work were made and these are listed below. Given the current stock status there is no urgency to schedule another benchmark for this stock in the short term.

## Time-varying M

The stock shows very strong changes in weights-at-age over time (they can change by a factor of up to 2). This is likely to affect the natural mortality. Further information to support this would be very useful for future benchmarks.

## Dome-shaped selectivity surveys

There are very little data to inform the question whether survey catchability is flat-topped or dome-shaped. At the moment, the highly truncated age structure means that this makes little difference in the model outputs. However if the stock recovers and more older fish appear then this will need to be revisited.

## FSP survey

The FSP survey potentially has useful information on the older fish (even though the survey is discontinued). Including the survey in the final assessment run resulted in many of the retrospective runs to fail to converge. It appears therefore that it causes the model to be unstable and was omitted from the final run. For future benchmarks, it may be useful to investigate why this survey makes the model unstable.

### 38.11 Management considerations

Discarding in the Nephrops fishery is the main management issue. Despite the implementation of several technical measures, which experimentally reduce whiting catches, as part of the cod longterm management plan the discards estimates still remain c .700 t . Given the continued high discards and low TAC, this stock could become a major 'choke species' for the 7.a Nephrops fishery in the context of the landing obligation.

Effort limitations are in force within the Irish Sea as a result of the cod long-term management plan. These effort limitations have not significant reduced mortality on whiting.

Whiting has a low market value, which is likely to contribute to discarding rates.

Technical measures applied to this stock include a minimum conservation reference size ( $\geq 27 \mathrm{~cm}$ ), whiting now mature well below this MCRS.

### 38.12 References

ICES. 2012. ICES Implementation of Advice for Data-limited Stocks in 2012 in its 2012 Advice. ICES CM 2012/ACOM 68.42 pp.

ICES. 2016. Report of the Workshop to consider MSY proxies for stocks in ICES category 3 and 4 stocks in Western Waters (WKProxy), 3-6 November 2015, ICES Headquarters, Copenhagen, Denmark. ICES CM 2015/ACOM:61. 159 pp.

ICES. 2017 :Report of the Benchmark Workshop on the Irish Sea Ecosystem (WKIrish3), 30 January-3 February 2017, Galway, Ireland, ICES CM 2017/BSG:01.

ICES. 2017. Report of the Second Workshop on the Impact of Ecosystem and Environmental Drivers on Irish Sea Fisheries Management (WKIrish2), 26-29 September 2016, Belfast, Northern Ireland, ICES CM 2016/BSG:02.

Table 40.1. Official Landings ( $\mathbf{t}$ ) of whiting in Division 7.a, 1988-2018, as reported to ICES.

| $\begin{aligned} & \frac{1}{\pi} \\ & \underset{\sim}{\sim} \end{aligned}$ |  | $\begin{aligned} & \text { 쁠 } \\ & \text { 든 } \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \text { 플 } \end{aligned}$ |  |  |  |  |  | 득 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 90 | 1,063 | 4,394 |  | 5,823 |  | 15 | 107 |  | 11,492 |
| 1989 | 92 | 533 | 3,871 |  | 6,652 |  | 26 | 154 |  | 11,328 |
| 1990 | 142 | 528 | 2,000 |  | 5,202 |  | 75 | 236 |  | 8,183 |
| 1991 | 53 | 611 | 2,200 |  | 4,250 |  | 74 | 223 |  | 7,411 |
| 1992 | 78 | 509 | 2,100 |  | 4,089 |  | 44 | 274 |  | 7,094 |
| 1993 | 50 | 255 | 1,440 |  | 3,859 |  | 55 | 318 |  | 5,977 |
| 1994 | 80 | 163 | 1,418 |  | 3,724 |  | 44 | 208 |  | 5,637 |
| 1995 | 92 | 169 | 1,840 |  | 3,125 |  | 41 | 198 |  | 5,465 |
| 1996 | 80 | 78 | 1,773 | 17 | 3,557 |  | 28 | 48 |  | 5,581 |
| 1997 | 47 | 86 | 1,119 | 14 | 3,152 |  | 24 | 30 |  | 4,472 |
| 1998 | 52 | 81 | 1,260 | 7 | 1,900 |  | 33 | 22 |  | 3,355 |
| 1999 | 46 | 150 | 509 | 6 | 1,229 |  | 5 | 44 |  | 1,989 |
| 2000 | 30 | 59 | 353 | 1 | 670 |  | 2 | 15 |  | 1,130 |
| 2001 | 27 | 25 | 482 |  | 506 |  | 1 | 25 |  | 1,066 |
| 2002 | 22 | 33 | 347 |  | 284 |  | 1 | 27 |  | 714 |
| 2003 | 13 | 29 | 265 |  | 130 | 85 | 1 | 31 |  | 554 |
| 2004 | 11 | 8 | 96 |  | 82 |  | 1 | 6 |  | 204 |
| 2005 | 10 | 13 | 94 |  | 47 |  |  | <0.5 |  | 164 |
| 2006 | 4 | 4 | 55 |  | 22 |  |  | <0.5 |  | 85 |
| 2007 | 3 | 3 | 187 |  | 3 |  | 1 | <0.5 |  | 197 |
| 2008 | 2 | 2 | 68 |  | 11 |  | 1 |  |  | 84 |
| 2009 | 2 |  | 78 |  | 20 |  |  |  |  | 100 |
| 2010 | 5 | 3 | 97 |  | 16 |  | <0.5 |  |  | 121 |
| 2011 | 4 | 3 | 95 |  | 16 |  | <0.5 |  |  | 118 |
| 2012 | 5 | 1 | 58 |  | 10 |  |  | 1 | 11 | 86 |
| 2013 | 2 | <0.5 | 44 |  |  |  | <0.1 | 2 | 20 | 68 |
| 2014 | 2 | <0.5 | 60 |  | 11 |  | <0.1 |  |  | 73 |
| 2,015 | 1 | <0.5 | 49 |  | 8 |  |  |  |  | 59 |
| 2,016 | 1 | <0.5 | 44 |  | 5 |  | <0.1 |  |  | 50 |
| 2,017 | 2 | <0.5 | 32 |  | 17 |  | <0.1 |  |  | 50 |
| 2018* | 1 |  | 44 |  | 19 |  | <0.5 |  |  | 63 |

[^26]Table 40.2. ICES estimates of discards, landings and catch of whiting in Division 7.a, 1988-2018.

| Year | Discards by Country/Fleet |  |  |  |  | Discards | Landings | Catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Nephrops fishery ${ }^{b}$ | IR-OTB fleet ${ }^{\text {ce }}$ | NI Nephrops fishery ${ }^{\text {d }}$ | Belgium | UK (E\&W) <br> fleet |  |  |  |
| 1988 | 1,611 |  |  |  |  | 1,611 | 10,245 | 11,856 |
| 1989 | 2,103 |  |  |  |  | 2,103 | 11,305 | 13,408 |
| 1990 | 2,444 |  |  |  |  | 2,444 | 8,212 | 10,656 |
| 1991 | 2,598 |  |  |  |  | 2,598 | 7,348 | 9,946 |
| 1992 | 4,203 |  |  |  |  | 4,203 | 8,588 | 12,791 |
| 1993 | 2,707 |  |  |  |  | 2,707 | 6,523 | 9,230 |
| 1994 | 1,173 |  |  |  |  | 1,173 | 6,763 | 7,936 |
| 1995 | 2,151 |  |  |  |  | 2,151 | 4,893 | 7,044 |
| 1996 | 3,631 |  |  |  |  | 3,631 | 4,335 | 7,966 |
| 1997 | 1,928 |  |  |  |  | 1,928 | 2,277 | 4,205 |
| 1998 | 1,304 |  |  |  |  | 1,304 | 2,229 | 3,533 |
| 1999 | 1,092 |  |  |  |  | 1,092 | 1,670 | 2,762 |
| 2000 | 2,118 |  |  |  |  | 2,118 | 762 | 2,880 |
| 2001 | 1,012 |  |  |  |  | 1,012 | 733 | 1,745 |
| 2002 | 740 |  |  |  |  | 740 | 747 | 1,487 |
| 2003 |  | 480 |  |  |  | 480 | 517 | 996 |
| 2004 |  | 905 |  |  |  | 905 | 133 | 1,038 |
| 2005 |  | 272 |  |  |  | 272 | 125 | 397 |
| 2006 |  | 1,580 | 193 |  |  | 1,773 | 64 | 1,837 |
| 2007 |  | 725 | 787 |  |  | 1,512 | 35 | 1,547 |
| 2008 |  | 693 | 476 |  |  | 1,169 | 37 | 1,206 |
| 2009 |  | 688 | 633 |  |  | 1,321 | 39 | 1,360 |
| 2010 |  | 240 | 914 |  |  | 1,154 | 30 | 1,184 |
| 2011 |  | 330 | 616 |  |  | 946 | 31 | 977 |
| 2012 |  | 257 | 1,065 | 17 | 1 | 1,339 | 60 | 1,399 |
| 2013 |  | 95 | 833 | 17 | 3 | 948 | 33 | 981 |
| 2014 |  | 263 | 1,645 | 15 | 28 | 1,951 | 23 | 1,974 |
| 2015 |  | 438 | 1,074 | 9 | 1 | 1,521 | 28 | 1,549 |
| 2016 |  | 173 | 589 |  | 3 | 765 | 15 | 780 |
| 2017 |  | 122 | 544 |  | 1 | 667 | 36 | 703 |
| 2018* |  | 98 | 754 |  | <0.5 | 853 | 46 | 899 |

b Based on UK(N.Ireland) and Ireland data.
c Based on data from Ireland.
d Based on data from Northern Ireland.

* Preliminary (and rounded).
e Raised using Days

Table 40.3. Whiting landings taken or reported in ICES rectangles $33 \mathrm{E} 2,33 \mathrm{E} 3$ and 33 E 4 have been reassigned to the 7.e-k whiting stock since 2003.

| Year | Official landings | ICES landings | ICES Discards | ICES catch | Landings taken or reported in rectangles 33E2 and 33E3 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 11,492 | 10,245 | 1,611 | 11,856 |  |
| 1989 | 11,328 | 11,305 | 2,103 | 13,408 |  |
| 1990 | 8,183 | 8,212 | 2,444 | 10,656 |  |
| 1991 | 7,411 | 7,348 | 2,598 | 9,946 |  |
| 1992 | 7,094 | 8,588 | 4,203 | 12,791 |  |
| 1993 | 5,977 | 6,523 | 2,707 | 9,230 |  |
| 1994 | 5,637 | 6,763 | 1,173 | 7,936 |  |
| 1995 | 5,465 | 4,893 | 2,151 | 7,044 |  |
| 1996 | 5,581 | 4,335 | 3,631 | 7,966 |  |
| 1997 | 4,472 | 2,277 | 1,928 | 4,205 |  |
| 1998 | 3,355 | 2,229 | 1,304 | 3,533 |  |
| 1999 | 1,989 | 1,670 | 1,092 | 2,762 |  |
| 2000 | 1,130 | 762 | 2,118 | 2,880 |  |
| 2001 | 1,066 | 733 | 1,012 | 1,745 |  |
| 2002 | 714 | 747 | 740 | 1,487 |  |
| 2003 | 554 | 517 | 480 | 996 | 159 |
| 2004 | 204 | 133 | 905 | 1,038 | 51 |
| 2005 | 164 | 125 | 272 | 397 | 33 |
| 2006 | 85 | 64 | 1,773 | 1,837 | 22 |
| 2007 | 197 | 35 | 1,512 | 1,547 | 161 |
| 2008 | 84 | 37 | 1,169 | 1,206 | 44 |
| 2009 | 100 | 39 | 1,321 | 1,360 | 63 |
| 2010 | 121 | 30 | 1,154 | 1,184 | 91 |
| 2011 | 118 | 31 | 946 | 977 | 75 |
| 2012 | 86 | 60 | 1,339 | 1,399 | 43 |
| 2013 | 68 | 33 | 948 | 981 | 33 |
| 2014 | 73 | 23 | 1,951 | 1,974 | 50 |
| 2015 | 59 | 28 | 1,521 | 1,549 | 34 |
| 2016 | 50 | 15 | 765 | 780 | 40 |
| 2017 | 50 | 36 | 667 | 703 | 20 |
| 2018 | 63 | 46 | 853 | 899 | 18 |

Table 40.4. Whiting7.a. Landings numbers-at-age.

|  | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0 | 14520 | 21811 | 6468 | 2548 | 350 | 0 |
| 1981 | 0 | 11203 | 29011 | 16004 | 2596 | 821 | 0 |
| 1982 | 41 | 5427 | 18098 | 19340 | 6108 | 813 | 0 |
| 1983 | 0 | 4886 | 9943 | 9100 | 4530 | 1165 | 321 |
| 1984 | 0 | 18254 | 12683 | 5257 | 2571 | 1045 | 402 |
| 1985 | 0 | 15540 | 35324 | 8687 | 996 | 0 | 675 |
| 1986 | 0 | 6306 | 16839 | 10809 | 1877 | 285 | 0 |
| 1987 | 0 | 10149 | 21563 | 6968 | 1943 | 242 | 0 |
| 1988 | 0 | 6983 | 25768 | 6989 | 1513 | 396 | 0 |
| 1989 | 0 | 11645 | 14029 | 13011 | 3645 | 490 | 0 |
| 1990 | 0 | 9502 | 17604 | 4734 | 1477 | 318 | 0 |
| 1991 | 102 | 7426 | 18406 | 5829 | 993 | 0 | 311 |
| 1992 | 0 | 8380 | 21907 | 7959 | 1374 | 462 | 0 |
| 1993 | 38 | 2742 | 21468 | 7327 | 932 | 0 | 135 |
| 1994 | 0 | 3245 | 6983 | 18509 | 1801 | 208 | 0 |
| 1995 | 0 | 1124 | 10095 | 3020 | 4444 | 233 | 0 |
| 1996 | 129 | 1652 | 6162 | 7432 | 1263 | 1082 | 135 |
| 1997 | 0 | 610 | 4239 | 2567 | 1795 | 87 | 79 |
| 1998 | 0 | 329 | 3287 | 4727 | 888 | 261 | 95 |
| 1999 | 1 | 341 | 2806 | 2607 | 741 | 160 | 119 |
| 2000 | 0 | 319 | 1364 | 1002 | 299 | 115 | 15 |
| 2001 | 0 | 111 | 1189 | 1006 | 171 | 53 | 20 |
| 2002 | 0 | 67 | 748 | 1480 | 376 | 48 | 41 |
| 2003 | 0 | 89 | 1051 | 606 | 199 | 0 | 0 |
| 2004 | 0 | 0 | 17 | 117 | 150 | 17 | 0 |
| 2005 | 0 | 0 | 101 | 216 | 95 | 21 | 3 |
| 2006 | 0 | 34 | 41 | 88 | 39 | 9 | 1 |
| 2007 | 0 | 24 | 41 | 32 | 10 | 3 | 0 |
| 2008 | 0 | 38 | 66 | 25 | 5 | 1 | 0 |
| 2009 | 0 | 65 | 44 | 22 | 4 | 1 | 0 |
| 2010 | 0 | 18 | 83 | 11 | 3 | 0 | 0 |
| 2011 | 0 | 1 | 17 | 59 | 15 | 3 | 0 |
| 2012 | 0 | 4 | 29 | 80 | 60 | 9 | 1 |
| 2013 | 8 | 81 | 36 | 20 | 5 | 1 | 1 |
| 2014 | 0 | 2 | 25 | 24 | 11 | 1 | 1 |
| 2015 | 0 | 2 | 25 | 24 | 11 | 1 | 1 |
| 2016 | 0 | 0 | 6 | 21 | 10 | 3 | 0 |
| 2017 | 0 | 0 | 9 | 50 | 43 | 5 | 1 |
| 2018 | 0 | 1 | 14 | 70 | 38 | 19 | 2 |

Table 40.5. Whiting7.a. Discards numbers-at-age.

|  | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 12786 | 32318 | 6888 | 65 | 26 | 0 | 0 |
| 1981 | 9865 | 24935 | 9162 | 162 | 26 | 0 | 0 |
| 1982 | 4047 | 8489 | 560 | 19 | 0 | 0 | 0 |
| 1983 | 23847 | 7328 | 2036 | 9 | 0 | 0 | 0 |
| 1984 | 26394 | 33900 | 1568 | 11 | 0 | 0 | 0 |
| 1985 | 12380 | 26461 | 1859 | 9 | 0 | 0 | 0 |
| 1986 | 28364 | 21111 | 1464 | 33 | 0 | 0 | 0 |
| 1987 | 16594 | 40598 | 1875 | 0 | 0 | 0 | 0 |
| 1988 | 6922 | 17958 | 1940 | 0 | 0 | 0 | 0 |
| 1989 | 17247 | 20701 | 2476 | 26 | 0 | 0 | 0 |
| 1990 | 4216 | 31810 | 3353 | 72 | 0 | 0 | 0 |
| 1991 | 20349 | 29334 | 3823 | 146 | 1 | 0 | 0 |
| 1992 | 1497 | 61451 | 10404 | 97 | 0 | 0 | 0 |
| 1993 | 12639 | 13979 | 17707 | 426 | 5 | 0 | 0 |
| 1994 | 3731 | 12063 | 1812 | 1702 | 29 | 0 | 0 |
| 1995 | 7118 | 17613 | 7015 | 492 | 234 | 0 | 0 |
| 1996 | 12732 | 39647 | 8168 | 1976 | 81 | 0 | 0 |
| 1997 | 8163 | 25497 | 5352 | 689 | 141 | 0 | 0 |
| 1998 | 6096 | 27131 | 2293 | 550 | 44 | 0 | 0 |
| 1999 | 20851 | 7677 | 2117 | 228 | 34 | 2 | 2 |
| 2000 | 7321 | 38922 | 4395 | 564 | 55 | 1 | 10 |
| 2001 | 16940 | 12631 | 3150 | 102 | 10 | 0 | 0 |
| 2002 | 8538 | 13412 | 1588 | 231 | 33 | 0 | 1 |
| 2003 | 12389 | 4595 | 201 | 0 | 0 | 0 | 0 |
| 2004 | 19699 | 14938 | 345 | 59 | 0 | 0 | 0 |
| 2005 | 643 | 5797 | 346 | 16 | 3 | 0 | 0 |
| 2006 | 15764 | 20590 | 613 | 21 | 0 | 0 | 0 |
| 2007 | 17436 | 24319 | 747 | 50 | 0 | 0 | 0 |
| 2008 | 10645 | 19994 | 676 | 16 | 0 | 0 | 0 |
| 2009 | 6622 | 27448 | 1176 | 0 | 0 | 0 | 0 |
| 2010 | 3946 | 15102 | 2810 | 64 | 1 | 0 | 0 |
| 2011 | 25982 | 8197 | 658 | 314 | 0 | 0 | 0 |
| 2012 | 6637 | 31020 | 790 | 37 | 1 | 3 | 0 |
| 2013 | 8493 | 11945 | 613 | 4 | 0 | 0 | 0 |
| 2014 | 13467 | 27553 | 2425 | 259 | 10 | 0 | 0 |
| 2015 | 3883 | 23595 | 2603 | 223 | 1 | 0 | 0 |
| 2016 | 4509 | 5780 | 4804 | 294 | 15 | 0 | 0 |
| 2017 | 3559 | 5870 | 4385 | 240 | 14 | 0 | 0 |
| 2018 | 6523 | 7386 | 2557 | 614 | 92 | 10 | 0 |

Table 40.6. Whiting in 7.a. Survey data available to WGCSE 2019.

NIGFS-WIBTS-Q1: Northern Ireland March Groundfish Survey

| 1992 | 2018 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0.21 | 0.25 |  |  |  |  |
| 1 | 6 |  |  |  |  |  |  |
| 1 | 665.6 | 710.3 | 81.2 | 11.7 | 4.3 | 0.8 | 1993 |
| 1 | 1804.6 | 262.1 | 299.2 | 44.7 | 11.9 | 8.1 | 1994 |
| 1 | 1688.9 | 635.7 | 174.2 | 88.4 | 22.0 | 6.3 | 1995 |
| 1 | 1468.4 | 334.0 | 213.0 | 35.1 | 37.2 | 5.4 | 1996 |
| 1 | 1406.1 | 1536.4 | 156.0 | 52.8 | 4.5 | 13.7 | 1997 |
| 1 | 1485.0 | 754.4 | 415.4 | 29.7 | 7.4 | 1.8 | 1998 |
| 1 | 1369.4 | 373.2 | 111.2 | 41.5 | 3.7 | 1.0 | 1999 |
| 1 | 2302.4 | 410.9 | 181.8 | 26.6 | 3.7 | 0.0 | 2000 |
| 1 | 1065.7 | 696.5 | 124.6 | 13.7 | 5.9 | 2.7 | 2001 |
| 1 | 2307.7 | 686.7 | 175.3 | 52.9 | 11.2 | 1.4 | 2002 |
| 1 | 1495.1 | 905.2 | 130.2 | 10.9 | 1.6 | 0.1 | 2003 |
| 1 | 1609.8 | 231.7 | 61.4 | 2.7 | 1.3 | 0.2 | 2004 |
| 1 | 689.3 | 124.0 | 28.5 | 12.3 | 2.8 | 0.1 | 2005 |
| 1 | 959.8 | 235.6 | 30.3 | 6.0 | 0.1 | 0.1 | 2006 |
| 1 | 905.0 | 158.6 | 14.9 | 2.7 | 0.2 | 0.0 | 2007 |
| 1 | 756.7 | 347.0 | 45.0 | 2.8 | 0.3 | 0.4 | 2008 |
| 1 | 1062.3 | 281.1 | 36.3 | 1.8 | 0.2 | 0.1 | 2009 |
| 1 | 739.4 | 545.8 | 51.6 | 4.7 | 6.4 | 0.0 | 2010 |
| 1 | 586.4 | 156.5 | 36.0 | 3.9 | 0.6 | 0.0 | 2011 |
| 1 | 972.2 | 354.4 | 42.3 | 5.9 | 1.2 | 0.0 | 2012 |
| 1 | 629.6 | 649.3 | 66.7 | 3.5 | 0.5 | 0.0 | 2013 |
| 1 | 922.1 | 367.6 | 67.0 | 4.3 | 0.2 | 0.1 | 2014 |
| 1 | 2797.3 | 469.3 | 18.8 | 2.3 | 0.0 | 0.0 | 2015 |
| 1 | 1409.1 | 924.8 |  | 38.7 | 1.5 | 0.1 | 2016 |
| 1 | 888.1 | 831.8 | 142.2 | 11.2 | 0.7 | 0.1 | 2017 |
| 1 | 431.4 | 296.8 | 119.4 | 17.9 | 2.3 | 0.0 | 2018 |

Table 40.6. Continued. Whiting in 7.a. Survey data available to WGCSE 2019.

NIGFS-WIBTS-Q4: Northern Ireland October Groundfish Survey

| 1993 | 2018 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0.83 | 0.88 |  |  |  |  |  |
| 0 | 6 |  |  |  |  |  |  |  |
| 1 | 714.0 | 1040.5 | 475.9 | 67.5 | 8.2 | 3.1 | 0.3 | 1993 |
| 1 | 1113.1 | 1320.0 | 208.6 | 150.7 | 33.9 | 2.3 | 0.5 | 1994 |
| 1 | 3124.4 | 477.3 | 166.5 | 30.6 | 35.6 | 5.4 | 1.2 | 1995 |
| 1 | 2306.2 | 591.2 | 134.4 | 52.4 | 10.5 | 7.0 | 1.3 | 1996 |
| 1 | 2626.5 | 676.6 | 497.6 | 61.0 | 18.2 | 4.6 | 4.5 | 1997 |
| 1 | 2863.5 | 466.8 | 153.8 | 72.8 | 6.2 | 2.2 | 0.1 | 1998 |
| 1 | 2478.4 | 1079.7 | 192.0 | 51.7 | 43.3 | 3.7 | 1.8 | 1999 |
| 1 | 2374.3 | 1084.7 | 126.0 | 20.0 | 16.9 | 6.0 | 2.7 | 2000 |
| 1 | 6356.4 | 658.3 | 270.8 | 28.9 | 4.9 | 2.3 | 0.0 | 2001 |
| 1 | 2692.4 | 1322.5 | 268.3 | 41.6 | 4.5 | 1.2 | 0.0 | 2002 |
| 1 | 4431.0 | 1572.3 | 921.1 | 74.8 | 16.8 | 1.5 | 0.0 | 2003 |
| 1 | 4457.1 | 699.6 | 268.3 | 113.8 | 4.4 | 1.9 | 0.0 | 2004 |
| 1 | 2377.2 | 487.8 | 183.3 | 15.8 | 1.5 | 0.4 | 0.0 | 2005 |
| 1 | 2849.2 | 144.8 | 46.8 | 7.9 | 1.8 | 0.0 | 0.0 | 2006 |
| 1 | 2163.1 | 957.6 | 149.1 | 16.7 | 4.8 | 4.3 | 0.2 | 2007 |
| 1 | 4884.6 | 1312.6 | 114.3 | 3.8 | 0.2 | 0.0 | 0.0 | 2008 |
| 1 | 2246.5 | 510.8 | 71.7 | 7.5 | 1.6 | 0.0 | 0.2 | 2009 |
| 1 | 2274.4 | 312.1 | 259.6 | 8.2 | 0.7 | 0.2 | 0.0 | 2010 |
| 1 | 3534.1 | 348.4 | 139.7 | 26.3 | 3.5 | 0.9 | 0.0 | 2011 |
| 1 | 1330.9 | 402.5 | 134.7 | 19.5 | 6.2 | 0.1 | 0.0 | 2012 |
| 1 | 7135.8 | 354.7 | 155.9 | 31.1 | 1.5 | 0.5 | 0.9 | 2013 |
| 1 | 4504.0 | 507.7 | 135.5 | 8.8 | 0.7 | 0.0 | 0.0 | 2014 |
| 1 | 2802.4 | 891.0 | 115.2 | 6.3 | 0.7 | 0.0 | 0.0 | 2015 |
| 1 | 2718.7 | 859.3 | 203.5 | 31.7 | 3.5 | 0.4 | 0 | 2016 |
| 1 | 3011.1 | 714.1 | 368.4 | 78.4 | 4.2 | 0.0 | 0.1 | 2017 |
| 1 | 4424.7 | 897.5 | 367.6 | 23.4 | 8.3 | 0.2 | 0.04 | 2018 |

Table 40.6. Continued. Whiting in 7.a. Survey data available to WGCSE 2019.

| Effort and numbers-at-age (per km towed) |  |  |  |
| :---: | :---: | :---: | :---: |
| 1988 | 2018 |  |  |
| 1 | 1 | 0.75 | 0.79 |
| 0 | 1 |  |  |
| 1 | 96 | 26 | 1988 |
| 1 | 93 | 21 | 1989 |
| 1 | 99 | 33 | 1990 |
| 1 | 216 | 25 | 1991 |
| 1 | 405 | 206 | 1992 |
| 1 | 253 | 95 | 1993 |
| 1 | 205 | 125 | 1994 |
| 1 | 1949 | 87 | 1995 |
| 1 | 169 | 194 | 1996 |
| 1 | 409 | 254 | 1997 |
| 1 | 893 | 199 | 1998 |
| 1 | 550 | 137 | 1999 |
| 1 | 320 | 122 | 2000 |
| 1 | 585 | 195 | 2001 |
| 1 | 280 | 96 | 2002 |
| 1 | 456 | 229 | 2003 |
| 1 | 917 | 330 | 2004 |
| 1 | 849 | 294 | 2005 |
| 1 | 1010 | 228 | 2006 |
| 1 | 339 | 89 | 2007 |
| 1 | 780 | 72 | 2008 |
| 1 | 389 | 371 | 2009 |
| 1 | 324 | 33 | 2010 |
| 1 | 1002 | 341 | 2011 |
| 1 | 442 | 426 | 2012 |
| 1 | 1535 | 228 | 2013 |
| 1 | 261 | 113 | 2014 |
| 1 | 211 | 112 | 2015 |
| 1 | 666 | 213 | 2016 |
| 1 | 489 | 230 | 2017 |
| 1 | 662 | 380 | 2018 |

Table 40.6. Continued. Whiting in 7.a. Survey data available to WGCSE 2019.
NIMIK : Northern Ireland MIK Net Survey

| 1994 | 2018 |  |  |
| :--- | :--- | :--- | :--- |
| 1 | 1 | 0.46 | 0.50 |
| 0 | 0 |  |  |
| 1 | 778 | 1994 |  |
| 1 | 225 | 1995 |  |
| 1 | 397 | 1996 |  |
| 1 | 205 | 1997 |  |
| 1 | 59 | 1998 |  |
| 1 | 91 | 1999 |  |
| 1 | 40 | 2000 |  |
| 1 | 167 | 2001 |  |
| 1 | 19 | 2002 |  |
| 1 | 148 | 2003 |  |
| 1 | 101 | 2004 |  |
| 1 | 135 | 2005 |  |
| 1 | 118 | 2006 |  |
| 1 | 82 | 2007 |  |
| 1 | 99 | 2008 |  |
| 1 | 173 | 2009 |  |
| 1 | 78 | 2010 |  |
| 1 | 122.2 | 2011 |  |
| 1 | 123.9 | 2012 |  |
| 1 | 197.6 | 2013 |  |
| 1 | 54.9 | 2014 |  |
| 1 | 59.5 | 2015 |  |
| 1 | 6.7 | 2016 |  |
| 1 | 175.5 | 2017 |  |
| 1 | 90.7 | 2018 |  |

Eastern Irish Sea FSP: Isadale 2005-2013: Numbers of fish per hour towed

| 2005 | 2013 |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 1 | 0.2 | 0.2 |  |  |  |  |
| 1 | 6.0 |  |  |  |  |  |  |
| 1 | 0.2 | 11.1 | 21.1 | 5.3 | 1.0 | 0.0 | 0.7 |
| 1 | 8.7 | 46.7 | 15.2 | 1.9 | 0.5 | 0.0 | 0.0 |
| 1 | 4.2 | 10.8 | 5.6 | 1.0 | 0.3 | 0.0 | 0.0 |
| 1 | 3.7 | 10.3 | 8.6 | 2.0 | 0.4 | 0.3 | 0.0 |
| 1 | 27.3 | 84.9 | 48.7 | 3.6 | 0.3 | 0.0 | 0.0 |
| 1 | 4.5 | 57.9 | 43.5 | 5.0 | 0.2 | 0.1 | 0.0 |
| 1 | 2.2 | 8.4 | 31.9 | 5.1 | 1.0 | 0.0 | 0.0 |
| 1 | 5.2 | 80.9 | 29.8 | 22.1 | 1.2 | 0.1 | 0.0 |
| 1 | 4.2 | 47.4 | 26.4 | 3.1 | 1.7 | 0.0 | 0.0 |

Table 40.7. Whiting7.a. ASAP input data.

```
# ASAP VERSION 3.0
#
# Number of Years
39
# First Year
1 9 8 0
# Number of Ages
7
# Number of Fleets
1
# Number of Selectivity Blocks
2
# Number of Available Indices
5
# Natural Mortality
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030}0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030}00.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
```



```
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030}00.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030}00.7180 0.6080 0.5540 0.5180 0.5180
```



```
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030}0.7180 0.6080 0.5540 0.5180 0.5180
```

```
1.0780 0.8030}00.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030}0.7180 0.6080 0.5540 0.5180 0.5180
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030}00.7180 0.6080 0.5540 0.5180 0.5180
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030}00.7180 0.6080 0.5540 0.5180 0.5180
1.0780}00.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030}00.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030 0.7180 0.6080 0.5540 0.5180 0.5180
1.0780 0.8030}00.7180 0.6080 0.5540 0.5180 0.5180
# Fecundity option
0
# Fraction of Year Prior to Spawning
0 . 0 0 0 0
# Maturity
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
```

```
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000}1.00001.000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
0.0000 0.0000 1.0000 1.0000 1.0000 1.0000 1.0000
```

\# Number of Weights-at-age Matrices
2
\# Weight Matrix 1
$\begin{array}{llllllllll}0.0000 & 0.1100 & 0.2350 & 0.3630 & 0.5290 & 0.6300 & 0.7720\end{array}$
$\begin{array}{llllllllll}0.0400 & 0.1180 & 0.2400 & 0.3640 & 0.5290 & 0.6300 & 0.8880\end{array}$
$\begin{array}{lllllll}0.0310 & 0.1350 & 0.2650 & 0.3650 & 0.5330 & 0.6300 & 0.7360\end{array}$
$0.0330 \quad 0.1460 \quad 0.2560 \quad 0.3970 \quad 0.4910 \quad 0.60500 .6550$
$0.0320 \quad 0.1250 \quad 0.2440 \quad 0.4030 \quad 0.5500 \quad 0.70000 .7450$
$0.02100 .10700 .2450 \quad 0.3330 \quad 0.47800 .56700 .6420$
$0.0250 \quad 0.1000 \quad 0.2170 \quad 0.3420 \quad 0.5120 \quad 0.70900 .9400$
$\begin{array}{llllllll}0.0240 & 0.1010 & 0.2170 & 0.3630 & 0.5350 & 0.7200 & 0.9330\end{array}$

```
0.0210 0.0880 0.2010 0.3300 0.5470 0.7630 1.0050
0.0260}00.1110 0.1930 0.2690 0.4330 0.6800 1.0790
0.0360}0.09400.2040 0.3100 0.4360 0.6760 0.8000
0.0310}00.0770 0.1940 0.2630 0.3520 0.4530 0.6920
0.0140 0.0630 0.1700 0.2720 0.3610 0.5130 1.0070
0.0290 0.0670 0.1420 0.2280 0.3310 0.4540 0.8920
0.0300 0.0740 0.1830 0.2210 0.3010 0.3780 0.4960
0.0310 0.0630 0.1790 0.2570 0.3260 0.5510 1.3200
0.0270 0.0570 0.1590 0.2300 0.2840 0.3640 0.7150
0.0260}00.0440 0.1530 0.2220 0.2870 0.3960 0.6790
0.0170}00.0350 0.1560 0.2280 0.2680 0.3500 0.4210
0.0280}00.0440 0.1610 0.2460 0.3240 0.3510 0.3250
0.0240}00.0380 0.1270 0.2180 0.2910 0.3470 0.3100
0.0170 0.0360 0.1320 0.3010 0.3380 0.5380}00.337
0.0160 0.0330 0.1240 0.2530 0.3390 0.4490 0.4250
0.0200 0.0480 0.2320 0.2950 0.2590 0.0000 0.0000
0.0170 0.0340 0.1310 0.3240 0.5090 0.4660 0.0000
0.0170}00.0370 0.1480 0.2630 0.3630 0.3600 0.3200
0.0170 0.0690 0.1520 0.2680 0.3610}00.3600 0.320
0.0230}00.0420 0.1220 0.2950 0.4340 0.6240 1.2600
0.0220}00.0440 0.1180 0.2620 0.3740 0.8340 1.3540
0.0230 0.0390 0.0940 0.3400 0.3230 0.5430 0.0000
0.0200 0.0480 0.1250 0.2560 0.4010 0.3750 0.0000
0.0180 0.0440 0.1040 0.1960 0.4050 0.4620 0.7990
0.0230}00.0350 0.1090 0.2750 0.3980 0.4100 0.3050
0.0300 0.0520 0.1120 0.2400 0.3460 0.2800 0.3800
0.0300 0.0420 0.1330}00.2260 0.4250 0.6590 1.0120
0.0220 0.0440 0.1270 0.2910 0.4480 0.2980}00.482
0.0220}00.0350 0.0850 0.1950 0.3410 0.4660 0.8820
0.0280 0.0320 0.0750 0.1980 0.3620 0.4320 0.5000
0.0210}00.0450 0.1040 0.1610 0.2400 0.3190 0.408
# Weight Matrix 2
0.0000 0.0733 0.1733 0.2992 0.4460 0.5795 0.7203
0.0000}00.0785 0.1797 0.3003 0.4468 0.5795 0.7143
0.0000 0.0840 0.1873 0.3110 0.4408 0.5760}00.694
0.0000 0.0850 0.1940 0.3210 0.4500 0.5813 0.6668
0.0000 0.0790 0.1918 0.3163 0.4473 0.5743 0.6628
```

```
0.0000 0.0697 0.1807 0.3038}00.4455 0.5825 0.6998
0.0000}00.0643 0.1685 0.2907 0.4338 0.5893 0.7485
0.0000}00.0598 0.1572 0.2857 0.4387 0.6195 0.8123
0.0000 0.0617 0.1500 0.2662 0.4250}00.6262 0.8682
0.0000 0.0607 0.1497 0.2533 0.3963 0.6057 0.8412
0.0000 0.0608 0.1473 0.2400 0.3550 0.5375 0.7817
0.0000 0.0545 0.1417 0.2393 0.3318}00.47720.718
0.0000 0.0480 0.1233 0.2218 0.3148 0.4282 0.7055
0.0000 0.0463 0.1170 0.2045 0.2927 0.3982 0.6358
0.0000 0.0462 0.1180}0.2002 0.2798 0.3960 0.6755
0.0000 0.0473 0.1208 0.2020 0.2695 0.3752 0.6523
0.0000 0.0420 0.1142 0.2050 0.2675 0.3703 0.6678
0.0000 0.0367 0.1053 0.1952 0.2580 0.3345 0.5210
0.0000 0.0322 0.1010 0.1940 0.2598}00.32270.422
0.0000 0.0313 0.0945 0.1937 0.2632 0.3212 0.3588
0.0000 0.0312 0.0895 0.2015 0.2742 0.3532 0.3367
0.0000 0.0293 0.0835 0.1987 0.2888 0.3812 0.3847
0.0000 0.0290 0.0992 0.2054 0.2847 0.4021 0.4114
0.0000}00.0281 0.1007 0.2267 0.3261 0.3847 0.4357
0.0000}00.0288 0.1045 0.2282 0.3338 0.3984 0.4062
0.0000 0.0323 0.0918 0.2277 0.3525 0.3862 0.3827
0.0000 0.0331 0.0939 0.2097 0.3355 0.4296 0.5145
0.0000 0.0352 0.0901 0.2082 0.3326 0.4961 0.7133
0.0000 0.0311 0.0815 0.2152 0.3261 0.5283 0.9183
0.0000 0.0331 0.0770 0.1989 0.3325 0.4804 0.9181
0.0000 0.0326 0.0756 0.1883 0.3311 0.4127 0.7840
0.0000 0.0313 0.0780 0.1750 0.3326 0.3957 0.5933
0.0000 0.0320 0.0753 0.1748 0.3127 0.3924 0.4550
0.0000}00.0334 0.0808 0.1777 0.3134 0.4162 0.4746
0.0000}00.0369 0.0836 0.1851 0.3267 0.4009 0.5369
0.0000 0.0339 0.0805 0.1806 0.3283 0.4403 0.6021
0.0000 0.0308 0.0680 0.1713 0.3104 0.4016 0.5479
0.0000 0.0306 0.0625 0.1401 0.2712 0.3946 0.5050
0.0000 0.0317 0.0615 0.1297 0.2489 0.3634 0.4623
```

\# Weight-at-age Pointers

```
1
1
2
2
# Selectivity Blocks
```

\# FLEET-1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
2
2
2
2
2
2
2
\# Selectivity Types
22
\# Selectivity Block Spec
\# Block 1

| 0.0000 | 1 | 0.0000 | 0.2500 |
| :---: | :---: | :---: | :---: |
| 0.5000 | 1 | 0.0000 | 0.2500 |
| 0.9000 | 1 | 0.0000 | 0.2500 |
| 1.0000 | -1 | 0.0000 | 0.2500 |
| 1.0000 | -1 | 0.0000 | 0.2500 |
| 1.0000 | -1 | 0.0000 | 0.2500 |
| 1.0000 | -1 | 0.0000 | 0.2500 |
| 3.0000 | 1 | 0.0000 | 1.0000 |
| 0.5000 | 1 | 0.0000 | 1.0000 |
| 0.0000 | 0 | 0.0000 | 0.0000 |
| 0.0000 | 0 | 0.0000 | 0.0000 |
| 0.0000 | 0 | 0.0000 | 0.0000 |
| 0.0000 | 0 | 0.0000 | 0.0000 |
| \# Block 2 |  |  |  |
| 0.2000 | 1 | 0.0000 | 0.5000 |
| 1.0000 | -1 | 0.0000 | 0.0000 |
| 1.0000 | -1 | 0.0000 | 0.0000 |
| 1.0000 | -1 | 0.0000 | 0.0000 |
| 1.0000 | -1 | 0.0000 | 0.0000 |
| 1.0000 | -1 | 0.0000 | 0.0000 |
| 1.0000 | -1 | 0.0000 | 0.0000 |
| 2.0000 | 1 | 0.0000 | 1.0000 |
| 0.5000 | 1 | 0.0000 | 1.0000 |
| 0.0000 | 0 | 0.0000 | 0.0000 |
| 0.0000 | 0 | 0.0000 | 0.0000 |


| 0.00000 | $0.0000 \quad 0.0$ | 000 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.00000 | 0.00000. | 000 |  |  |  |  |  |
| \# Fleet Selectivity Start Age |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| \# Fleet Selectivity End Age |  |  |  |  |  |  |  |
| 7 |  |  |  |  |  |  |  |
| \# Age Range Average F |  |  |  |  |  |  |  |
| 24 |  |  |  |  |  |  |  |
| \# Average F Report Option |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| \# Use Likelihood Constants |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| \# Release Mortality |  |  |  |  |  |  |  |
| 1.0000 |  |  |  |  |  |  |  |
| \# Catch at Age FLEET-1 |  |  |  |  |  |  |  |
| 12786.000 | 46838.000 | 28699.000 | 6533.000 | 2574.000 | 350.000 | 621.000 | 16737.000 |
| 9865.000 | 36138.000 | 38173.000 | 16166.000 | 2622.000 | 821.000 | 339.000 | 21331.000 |
| 4088.000 | 13916.000 | 18658.000 | 19359.000 | 6108.000 | 813.000 | 400.000 | 17969.000 |
| 23847.000 | 12214.000 | 11979.000 | 9109.000 | 4530.000 | 1165.000 | 321.000 | 12405.000 |
| 26394.000 | 52154.000 | 14250.000 | 5268.000 | 2571.000 | 1045.000 | 402.000 | 14999.000 |
| 12380.000 | 42001.000 | 37183.000 | 8696.000 | 996.000 | 675.000 | 372.000 | 18169.000 |
| 28364.000 | 27417.000 | 18303.000 | 10842.000 | 1877.000 | 285.000 | 270.000 | 12129.000 |
| 16594.000 | 50747.000 | 23438.000 | 6968.000 | 1943.000 | 242.000 | 111.000 | 14270.000 |
| 6922.000 | 24941.000 | 27708.000 | 6989.000 | 1513.000 | 396.000 | 197.000 | 11856.000 |
| 17247.000 | 32346.000 | 16505.000 | 13037.000 | 3645.000 | 490.000 | 177.000 | 13408.000 |
| 4216.000 | 41312.000 | 20957.000 | 4806.000 | 1477.000 | 318.000 | 128.000 | 10656.000 |
| 20451.000 | 36760.000 | 22229.000 | 5975.000 | 994.000 | 311.000 | 84.000 | 9946.000 |
| 1497.000 | 69831.000 | 32311.000 | 8056.000 | 1374.000 | 462.000 | 93.000 | 12791.000 |
| 12677.000 | 16721.000 | 39175.000 | 7753.000 | 937.000 | 135.000 | 27.000 | 9230.000 |
| 3731.000 | 15308.000 | 8795.000 | 20211.000 | 1830.000 | 208.000 | 50.000 | 7936.000 |
| 7118.000 | 18737.000 | 17110.000 | 3512.000 | 4678.000 | 233.000 | 21.000 | 7044.000 |
| 12861.000 | 41299.000 | 14330.000 | 9408.000 | 1344.000 | 1082.000 | 135.000 | 7966.000 |
| 8163.000 | 26107.000 | 9591.000 | 3256.000 | 1936.000 | 87.000 | 79.000 | 4205.000 |
| 6096.000 | 27460.000 | 5580.000 | 5277.000 | 932.000 | 261.000 | 95.000 | 3533.000 |
| 20852.000 | 8018.000 | 4923.000 | 2835.000 | 776.000 | 161.000 | 121.000 | 2762.000 |
| 7321.000 | 39242.000 | 5758.000 | 1566.000 | 354.000 | 115.000 | 25.000 | 2880.000 |
| 16940.000 | 12742.000 | 4338.000 | 1108.000 | 181.000 | 53.000 | 20.000 | 1745.000 |


| 8538.000 | 13480.000 | 2336.000 | 1710.000 | 408.000 | 48.000 | 42.000 | 1487.000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 12389.000 | 4685.000 | 1252.000 | 606.000 | 199.000 | 0.000 | 0.000 | 996.000 |
| 19699.000 | 14938.000 | 362.000 | 176.000 | 150.000 | 17.000 | 0.000 | 1038.000 |
| 643.000 | 5797.000 | 448.000 | 232.000 | 98.000 | 21.000 | 3.000 | 397.000 |
| 15764.000 | 20624.000 | 654.000 | 109.000 | 39.000 | 9.000 | 1.000 | 1837.000 |
| 17436.000 | 24343.000 | 787.000 | 82.000 | 10.000 | 3.000 | 0.000 | 1547.000 |
| 10645.000 | 20032.000 | 742.000 | 41.000 | 5.000 | 1.000 | 0.000 | 1206.000 |
| 6622.000 | 27513.000 | 1220.000 | 22.000 | 4.000 | 1.000 | 0.000 | 1360.000 |
| 3946.000 | 15120.000 | 2894.000 | 75.000 | 4.000 | 0.000 | 0.000 | 1184.000 |
| 25982.000 | 8198.000 | 675.000 | 373.000 | 15.000 | 3.000 | 0.000 | 977.000 |
| 6637.000 | 31023.000 | 819.000 | 116.000 | 61.000 | 12.000 | 1.000 | 1399.000 |
| 8501.000 | 12026.000 | 649.000 | 24.000 | 5.000 | 1.000 | 1.000 | 981.000 |
| 13467.000 | 27555.000 | 2450.000 | 284.000 | 21.000 | 1.000 | 1.000 | 1974.000 |
| 3883.000 | 23595.000 | 2613.000 | 267.000 | 15.000 | 1.000 | 1.000 | 1549.000 |
| 4509.000 | 5780.000 | 4809.000 | 315.000 | 25.000 | 3.000 | 0.000 | 780.000 |
| 3559.000 | 5871.000 | 4394.000 | 290.000 | 57.000 | 5.000 | 1.000 | 704.000 |
| 6523.000 | 7386.000 | 2571.000 | 684.000 | 129.000 | 29.000 | 2.000 | 899.000 |

\# Discards at Age FLEET-1

| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |


| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

\# Release Proportion at Age FLEET-1
$\begin{array}{lllllll}0.000000 & 0.000000 & 0.000000 & 0.000000 & 0.000000 & 0.000000 & 0.000000\end{array}$ $0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$ $0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$ $0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$ $0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$ $0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$

| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

$0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.0000000 .000000$
$0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$
$0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$
$\begin{array}{lllllll}0.000000 & 0.000000 & 0.000000 & 0.000000 & 0.000000 & 0.000000 & 0.000000\end{array}$
$0.000000 \quad 0.000000 \quad 0.0000000 .000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$
$0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$
$0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$
$0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000 \quad 0.000000$

| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |

\# Survey Specifications
\# Survey Units
22222
\# Survey Age ProportionUnits

22222
\# Survey Weight-at-age Matrix
11111
\# Survey Month
310595
\# Survey Choice
-1 -1 -1 -1 -1
\# Survey Selectivity Type

11112
\# Survey Start Age

```
21112
# Survey End Age
77127
# Estimate Survey Proportion at Age ?
11000
# Use Survey ?
11100
# Index Selectivity
# INDEX-1
0.000000 -1 0.000000 0.000000
0.500000 1 1 0.000000 0.500000
1.000000 -1 0.000000 0.000000
1.000000 -1 0.000000 0.000000
1.000000 -1 0.000000 0.000000
1.000000 -1 0.000000 0.000000
1.000000 -1 0.000000 0.000000
2.000000 1 0.000000 1.000000
0.500000 1 0.000000 1.000000
0.000000 0 0.000000 0.000000
0.000000 0 0.000000 0.000000
0.000000 0 0.000000 0.000000
0.000000 0 0.000000 0.000000
# INDEX-2
0.200000 1 0.000000 0.500000
0.500000 1 0.000000 0.500000
1.000000
1.000000
1.000000
1.000000
1.000000 -1 0.000000 0.000000
2.000000 1 0.000000 1.000000
0.500000 1 0.000000 1.000000
0.000000 0 0.000000 0.000000
0.000000 0 0.000000 0.000000
0.000000 0 0.000000 0.000000
0.000000 0}00.000000\quad0.00000
```

\# INDEX-3

| 1.000000 | -1 | 0.000000 | 0.000000 |
| :---: | :---: | :---: | :---: |
| 0.000000 | -1 | 0.000000 | 0.000000 |
| 0.000000 | -1 | 0.000000 | 0.000000 |
| 0.000000 | -1 | 0.000000 | 0.000000 |
| 0.000000 | -1 | 0.000000 | 0.000000 |
| 0.000000 | -1 | 0.000000 | 0.000000 |
| 0.000000 | -1 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| \# INDEX-4 |  |  |  |
| 1.000000 | -1 | 0.000000 | 1.000000 |
| 1.000000 | -1 | 0.000000 | 0.000000 |
| 1.000000 | -1 | 0.000000 | 0.000000 |
| 1.000000 | -1 | 0.000000 | 0.000000 |
| 1.000000 | -1 | 0.000000 | 0.000000 |
| 1.000000 | -1 | 0.000000 | 0.000000 |
| 1.000000 | -1 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |
| 0.000000 | 0 | 0.000000 | 0.000000 |

\# INDEX-5
$0.000000 \quad 0 \quad 0.000000 \quad 0.000000$
$0.000000 \quad 0 \quad 0.000000 \quad 0.000000$
$0.000000 \quad 0 \quad 0.000000 \quad 0.000000$
$0.000000 \quad 0 \quad 0.000000 \quad 0.000000$
$0.000000 \quad 0 \quad 0.000000 \quad 0.000000$
$0.000000 \quad 0 \quad 0.000000 \quad 0.000000$
$0.000000 \quad 0 \quad 0.0000000 .000000$
$4.000000 \quad 1 \quad 0.0000001 .000000$
$0.500000 \quad 1 \quad 0.0000001 .000000$

| 0.000 | 0000000 | 0.0000000 .0 | . 000000 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.000 | 0000000 | 0.0000000. | . 0000000 |  |  |  |  |  |  |  |
| 0.000 | 0000000 | 0.0000000. | . 0000000 |  |  |  |  |  |  |  |
| 0.000 | 0000000 | 0.0000000. | . 0000000 |  |  |  |  |  |  |  |
| \# Index Data |  |  |  |  |  |  |  |  |  |  |
| \# INDEX-1 |  |  |  |  |  |  |  |  |  |  |
| 1980 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1981 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1982 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1983 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1984 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1985 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1986 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1987 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1988 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1989 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1990 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1991 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1992 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1993 | 1474.000000 | 00.300000 | 0.000000 | 0.452000 | 0.482000 | 0.055000 | 0.008000 | 0.003000 | 0.001000 | 50.000000 |
| 1994 | 2431.000000 | 00.300000 | 0.000000 | 0.742000 | 0.108000 | 0.123000 | 0.018000 | 0.005000 | 0.003000 | 50.000000 |
| 1995 | 2615.000000 | 00.300000 | 0.000000 | 0.646000 | 0.243000 | 0.067000 | 0.034000 | 0.008000 | 0.002000 | 50.000000 |
| 1996 | 2093.000000 | 0.300000 | 0.000000 | 0.702000 | 0.160000 | 0.102000 | 0.017000 | 0.018000 | 0.003000 | 50.000000 |
| 1997 | 3169.000000 | 00.300000 | 0.000000 | 0.444000 | 0.485000 | 0.049000 | 0.017000 | 0.001000 | 0.004000 | 50.000000 |
| 1998 | 2694.000000 | 00.300000 | 0.000000 | 0.551000 | 0.280000 | 0.154000 | 0.011000 | 0.003000 | 0.001000 | 50.000000 |
| 1999 | 1900.000000 | 0.300000 | 0.000000 | 0.721000 | 0.196000 | 0.059000 | 0.022000 | 0.002000 | 0.001000 | 50.000000 |
| 2000 | 2925.000000 | 00.300000 | 0.000000 | 0.787000 | 0.140000 | 0.062000 | 0.009000 | 0.001000 | 0.000000 | 50.000000 |
| 2001 | 1909.000000 | 00.300000 | 0.000000 | 0.558000 | 0.365000 | 0.065000 | 0.007000 | 0.003000 | 0.001000 | 50.000000 |
| 2002 | 3235.000000 | 00.300000 | 0.000000 | 0.713000 | 0.212000 | 0.054000 | 0.016000 | 0.003000 | 0.000000 | 50.000000 |
| 2003 | 2543.000000 | 00.300000 | 0.000000 | 0.588000 | 0.356000 | 0.051000 | 0.004000 | 0.001000 | 0.000000 | 50.000000 |
| 2004 | 1907.000000 | 00.300000 | 0.000000 | 0.844000 | 0.121000 | 0.032000 | 0.001000 | 0.001000 | 0.000000 | 50.000000 |
| 2005 | 857.000000 | 0.300000 | 0.000000 | 0.804000 | 0.145000 | 0.033000 | 0.014000 | 0.003000 | 0.000000 | 50.000000 |
| 2006 | 1232.000000 | 0.300000 | 0.000000 | 0.779000 | 0.191000 | 0.025000 | 0.005000 | 0.000000 | 0.000000 | 50.000000 |
| 2007 | 1081.000000 | 0.300000 | 0.000000 | 0.837000 | 0.147000 | 0.014000 | 0.002000 | 0.000000 | 0.000000 | 50.000000 |
| 2008 | 1152.000000 | 0.300000 | 0.000000 | 0.657000 | 0.301000 | 0.039000 | 0.002000 | 0.000000 | 0.000000 | 50.000000 |
| 2009 | 1382.000000 | 00.300000 | 0.000000 | 0.769000 | 0.203000 | 0.026000 | 0.001000 | 0.000000 | 0.000000 | 50.000000 |
| 2010 | 1348.000000 | 00.300000 | 0.000000 | 0.549000 | 0.405000 | 0.038000 | 0.003000 | 0.005000 | 0.000000 | 50.000000 |


| 2011 | 783.000000 | 0.300000 | 0.000000 | 0.749000 | 0.200000 | 0.046000 | 0.005000 | 0.001000 | 0.000000 | 50.000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 1376.000000 | 0.300000 | 0.000000 | 0.707000 | 0.258000 | 0.031000 | 0.004000 | 0.001000 | 0.000000 | 50.000000 |
| 2013 | 1350.000000 | 0.300000 | 0.000000 | 0.466000 | 0.481000 | 0.049000 | 0.003000 | 0.000000 | 0.000000 | 50.000000 |
| 2014 | 1361.000000 | 0.300000 | 0.000000 | 0.677000 | 0.270000 | 0.049000 | 0.003000 | 0.000000 | 0.000000 | 50.000000 |
| 2015 | 3288.000000 | 0.300000 | 0.000000 | 0.851000 | 0.143000 | 0.006000 | 0.001000 | 0.000000 | 0.000000 | 50.000000 |
| 2016 | 2374.000000 | 0.300000 | 0.000000 | 0.594000 | 0.390000 | 0.000001 | 0.016000 | 0.001000 | 0.000000 | 50.000000 |
| 2017 | 1874.000000 | 0.300000 | 0.000000 | 0.474000 | 0.444000 | 0.076000 | 0.006000 | 0.001000 | 0.000000 | 50.000000 |
| 2018 | 868.000000 | 0.300000 | 0.000000 | 0.497000 | 0.342000 | 0.138000 | 0.021000 | 0.003000 | 0.000000 | 0.021000 |
| \# IND | EX-2 |  |  |  |  |  |  |  |  |  |
| 1980 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1981 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1982 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1983 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1984 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1985 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1986 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1987 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1988 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1989 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1990 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1991 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1992 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1993 | 2309.000000 | 0.300000 | 0.309000 | 0.451000 | 0.206000 | 0.029000 | 0.004000 | 0.001000 | 0.000000 | 50.000000 |
| 1994 | 2829.000000 | 0.300000 | 0.393000 | 0.467000 | 0.074000 | 0.053000 | 0.012000 | 0.001000 | 0.000000 | 50.000000 |
| 1995 | 3841.000000 | 0.300000 | 0.813000 | 0.124000 | 0.043000 | 0.008000 | 0.009000 | 0.001000 | 0.000000 | 50.000000 |
| 1996 | 3103.000000 | 0.300000 | 0.743000 | 0.191000 | 0.043000 | 0.017000 | 0.003000 | 0.002000 | 0.000000 | 50.000000 |
| 1997 | 3889.000000 | 0.300000 | 0.675000 | 0.174000 | 0.128000 | 0.016000 | 0.005000 | 0.001000 | 0.001000 | 50.000000 |
| 1998 | 3566.000000 | 0.300000 | 0.803000 | 0.131000 | 0.043000 | 0.020000 | 0.002000 | 0.001000 | 0.000000 | 50.000000 |
| 1999 | 3851.000000 | 0.300000 | 0.644000 | 0.280000 | 0.050000 | 0.013000 | 0.011000 | 0.001000 | 0.000000 | 50.000000 |
| 2000 | 3631.000000 | 0.300000 | 0.654000 | 0.299000 | 0.035000 | 0.006000 | 0.005000 | 0.002000 | 0.001000 | 50.000000 |
| 2001 | 7322.000000 | 0.300000 | 0.868000 | 0.090000 | 0.037000 | 0.004000 | 0.001000 | 0.000000 | 0.000000 | 50.000000 |
| 2002 | 4331.000000 | 0.300000 | 0.622000 | 00.305000 | 0.062000 | 0.010000 | 0.001000 | 0.000000 | 0.000000 | 50.000000 |
| 2003 | 7017.000000 | 0.300000 | 0.631000 | 0.224000 | 0.131000 | 0.011000 | 0.002000 | 0.000000 | 0.000000 | 50.000000 |
| 2004 | 5545.000000 | 0.300000 | 0.804000 | 0.126000 | 0.048000 | 0.021000 | 0.001000 | 0.000000 | 0.000000 | 50.000000 |
| 2005 | 3066.000000 | 0.300000 | 0.775000 | 0.159000 | 0.060000 | 0.005000 | 0.000000 | 0.000000 | 0.000000 | 50.000000 |
| 2006 | 3050.000000 | 0.300000 | 0.934000 | 0.047000 | 0.015000 | 0.003000 | 0.001000 | 0.000000 | 0.000000 | 50.000000 |
| 2007 | 3296.000000 | 0.300000 | 0.656000 | 0.291000 | 0.045000 | 0.005000 | 0.001000 | 0.001000 | 0.000000 | 50.000000 |


| 2008 | 6315.000000 | 0.300000 | 0.773000 | 0.208000 | 0.018000 | 0.001000 | 0.000000 | 0.000000 | 0.000000 | 50.000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 2838.000000 | 0.300000 | 0.791000 | 0.180000 | 0.025000 | 0.003000 | 0.001000 | 0.000000 | 0.000000 | 50.000000 |
| 2010 | 2855.000000 | 0.300000 | 0.797000 | 0.109000 | 0.091000 | 0.003000 | 0.000000 | 0.000000 | 0.000000 | 50.000000 |
| 2011 | 4053.000000 | 0.300000 | 0.872000 | 0.086000 | 0.034000 | 0.006000 | 0.001000 | 0.000000 | 0.000000 | 50.000000 |
| 2012 | 1894.000000 | 0.300000 | 0.703000 | 0.213000 | 0.071000 | 0.010000 | 0.003000 | 0.000000 | 0.000000 | 50.000000 |
| 2013 | 7680.000000 | 0.300000 | 0.929000 | 0.046000 | 0.020000 | 0.004000 | 0.000000 | 0.000000 | 0.000000 | 50.000000 |
| 2014 | 5157.000000 | 0.300000 | 0.873000 | 0.098000 | 0.026000 | 0.002000 | 0.000000 | 0.000000 | 0.000000 | 50.000000 |
| 2015 | 3816.000000 | 0.300000 | 0.734000 | 0.234000 | 0.030000 | 0.002000 | 0.000000 | 0.000000 | 0.000000 | 50.000000 |
| 2016 | 3817.000000 | 0.300000 | 0.712000 | 0.225000 | 0.053000 | 0.008000 | 0.000000 | 0.000000 | 0.000000 | 50.000000 |
| 2017 | 4176.000000 | 0.300000 | 0.721000 | 0.171000 | 0.088000 | 0.019000 | 0.000000 | 0.000000 | 0.000000 | 50.000000 |
| 2018 | 5713.000000 | 0.300000 | 0.774000 | 0.157000 | 0.000000 | 0.064000 | 0.000000 | 0.000000 | 0.000000 | 50.000000 |
| \# IND | X-3 |  |  |  |  |  |  |  |  |  |
| 1980 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1981 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1982 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1983 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1984 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1985 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1986 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1987 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1988 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1989 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1990 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1991 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1992 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1993 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1994 | 778.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1995 | 225.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1996 | 397.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1997 | 205.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1998 | 59.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1999 | 91.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2000 | 40.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2001 | 167.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2002 | 19.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2003 | 148.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2004 | 101.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |


| 2005 | 135.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 118.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2007 | 82.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2008 | 99.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2009 | 173.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2010 | 78.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2011 | 122.200000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2012 | 123.900000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2013 | 197.600000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2014 | 54.900000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2015 | 59.500000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2016 | 6.700000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2017 | 175.400000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2018 | 90.740000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| \# INDEX-4 |  |  |  |  |  |  |  |  |  |  |
| 1980 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1981 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1982 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1983 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1984 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1985 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1986 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1987 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1988 | 96.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1989 | 93.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1990 | 99.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1991 | 216.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1992 | 405.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1993 | 253.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1994 | 205.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1995 | 1949.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1996 | 169.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1997 | 409.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1998 | 893.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1999 | 550.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2000 | 320.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2001 | 585.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |


| 2002 | 280.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 456.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2004 | 917.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2005 | 849.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2006 | 1010.000000 | 0.500000 | 00.000000 | 0.000000 | 00.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2007 | 339.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2008 | 780.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2009 | 389.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2010 | 324.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2011 | 1002.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2012 | 442.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2013 | 1535.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2014 | 261.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2015 | 211.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2016 | 666.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2017 | 489.000000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 2018 | 661.750000 | 0.500000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| \# INDEX-5 |  |  |  |  |  |  |  |  |  |  |
| 1980 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1981 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1982 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1983 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1984 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1985 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1986 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1987 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1988 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1989 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1990 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1991 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1992 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1993 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1994 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1995 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1996 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1997 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1998 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |

```
1999 0.000000
```



```
2001}00.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 
```



```
2003}00.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000
2004 0.000000 0.000000 0.000000 0.000000
2005 38.660000 0.500000 0.000000 0.000000 0.000000
2006 72.953000 0.500000 0.000000 0.000000 0.000000
2007 21.870000 0.500000 0.000000 0.000000 0.000000
llllllllllll
2009 164.820000
2010}111.120000 0.500000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000
2011}48.600000 0.500000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000
2012 139.250000}00.500000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000
2013 82.850000 0.500000 0.000000 0.000000 0.000000}00.000000 0.000000 0.000000 0.000000 0.000000
2014}00.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000
2015}00.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000
2016}00.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000
2017 0.000000}00.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000
2018}00.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000 0.000000
# Phase Flags
#
1
#
1
#
3
\# Recruit CV
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
1.0000
\# Lambda Index
1.0000001 .0000001 .0000001 .0000001 .000000
\# Lambda Catch
1.000000
\# Lambda Discard
0.000000
\# Catch CV
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
0.200000
\# Discard CV
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
\# Catch Sample Size
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
10.000000
\# Discard Sample Size
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
0.000000
\# FMult Lambda
0.000000
\# FMult cv
0.900000
\# F dev Lambda
0.000000
\# F dev cv
0.900000
\#N 1st Year Lambda
0.000000
\#N 1st Year cv
0.9000
\#Recruit Lambda
0.100000
\# Lambda
\(\begin{array}{lllll}0.000000 & 0.000000 & 0.000000 & 0.000000 & 0.000000\end{array}\)
```

\#cv
0.900000 0.900000 0.900000 0.900000 0.900000

# Lambda

    0.000000 0.000000 0.000000 0.000000 0.000000
    \#cv
0.900000 0.900000 0.900000 0.900000 0.900000

# Lambda Steepness

    0 . 0 0 0 0 0 0
    
# cv Steepness

    0.900000
    
# Lambda Unexploited Srock

    0.000000
    \#cv
0.900000

# Stock at Age in 1st Year Option

1

# Initial Guess Stock at Age in 1st Year

1000000.000000 500000.000000 250000.000000 125000.000000 60000.000000 30000.000000 10000.000000

# Initial Guess

1.000000

# Initial Guess

1.000000e-003 1.000000e-003 1.000000e-003 1.000000e-003 1.000000e-003

# Stock Recruitment Option

O

# Initial Guess

1.000000e+003

# Initial Guess

1.000000e+000

# Initial Guess

    2.5000
    
# Ignore Guesses

0

# Projection

0

# 

1

# 

```

2019
\#
\(2019-13-99.0000 \quad 1.0000\)
\#MCMC
\#

0
\#

1
\#

0
\#

0
\#

0
\#
\(-1\)
\#

0
\#

0
\# Export R
1
\#
\(-23456\)
\# FINIS

Table 40.8. Whiting 7.a. Selectivity of the catches and indices.
\begin{tabular}{ccccc}
\hline Age & Catch & NI-Q1 & NI-Q4 & NI-MIK \\
\hline 0 & 0.120 & 0.000 & 0.664 & 1.000 \\
\hline 1 & 0.836 & 0.560 & 0.749 & 0.000 \\
\hline 2 & 0.995 & 1.000 & 1.000 & 0.000 \\
\hline 3 & 1.000 & 1.000 & 1.000 & 1.000 \\
\hline 5 & 1.000 & 1.000 & 1.000 & 0.000 \\
\hline 6 & & & 1.000 & 0.000 \\
\hline
\end{tabular}

Table 40.9. Whiting7.a Fishing mortality- (F) -at age.
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline & Age 0 & Age 1 & Age 2 & Age 3 & Age 4 & Age 5 & Age 6 \\
\hline 1980 & 0.027 & 0.177 & 0.438 & 0.521 & 0.532 & 0.533 & 0.533 \\
\hline 1981 & 0.032 & 0.208 & 0.516 & 0.613 & 0.626 & 0.627 & 0.627 \\
\hline 1982 & 0.035 & 0.23 & 0.57 & 0.678 & 0.692 & 0.693 & 0.694 \\
\hline 1983 & 0.036 & 0.233 & 0.579 & 0.689 & 0.703 & 0.705 & 0.705 \\
\hline 1984 & 0.044 & 0.286 & 0.71 & 0.844 & 0.861 & 0.863 & 0.863 \\
\hline 1985 & 0.053 & 0.347 & 0.862 & 1.025 & 1.046 & 1.048 & 1.048 \\
\hline 1986 & 0.041 & 0.27 & 0.671 & 0.798 & 0.815 & 0.816 & 0.817 \\
\hline 1987 & 0.044 & 0.287 & 0.712 & 0.847 & 0.864 & 0.866 & 0.866 \\
\hline 1988 & 0.037 & 0.243 & 0.603 & 0.717 & 0.732 & 0.734 & 0.734 \\
\hline 1989 & 0.052 & 0.341 & 0.846 & 1.006 & 1.027 & 1.03 & 1.03 \\
\hline 1990 & 0.044 & 0.289 & 0.718 & 0.854 & 0.872 & 0.874 & 0.874 \\
\hline 1991 & 0.044 & 0.29 & 0.72 & 0.857 & 0.874 & 0.876 & 0.877 \\
\hline 1992 & 0.072 & 0.47 & 1.167 & 1.388 & 1.417 & 1.42 & 1.42 \\
\hline 1993 & 0.057 & 0.373 & 0.926 & 1.102 & 1.124 & 1.127 & 1.127 \\
\hline 1994 & 0.059 & 0.385 & 0.957 & 1.137 & 1.161 & 1.164 & 1.164 \\
\hline 1995 & 0.103 & 0.718 & 0.855 & 0.859 & 0.859 & 0.859 & 0.859 \\
\hline 1996 & 0.113 & 0.79 & 0.94 & 0.945 & 0.945 & 0.945 & 0.945 \\
\hline 1997 & 0.098 & 0.684 & 0.813 & 0.817 & 0.817 & 0.817 & 0.817 \\
\hline 1998 & 0.137 & 0.954 & 1.135 & 1.141 & 1.141 & 1.141 & 1.141 \\
\hline 1999 & 0.108 & 0.757 & 0.9 & 0.905 & 0.905 & 0.905 & 0.905 \\
\hline 2000 & 0.142 & 0.991 & 1.179 & 1.185 & 1.185 & 1.185 & 1.185 \\
\hline 2001 & 0.115 & 0.801 & 0.953 & 0.958 & 0.958 & 0.958 & 0.958 \\
\hline 2002 & 0.157 & 1.096 & 1.303 & 1.31 & 1.31 & 1.31 & 1.31 \\
\hline 2003 & 0.077 & 0.541 & 0.644 & 0.647 & 0.647 & 0.647 & 0.647 \\
\hline 2004 & 0.215 & 1.504 & 1.789 & 1.798 & 1.798 & 1.798 & 1.798 \\
\hline 2005 & 0.056 & 0.394 & 0.469 & 0.472 & 0.472 & 0.472 & 0.472 \\
\hline 2006 & 0.192 & 1.344 & 1.598 & 1.607 & 1.607 & 1.607 & 1.607 \\
\hline 2007 & 0.152 & 1.059 & 1.26 & 1.266 & 1.266 & 1.266 & 1.266 \\
\hline 2008 & 0.122 & 0.855 & 1.018 & 1.023 & 1.023 & 1.023 & 1.023 \\
\hline 2009 & 0.139 & 0.972 & 1.157 & 1.163 & 1.163 & 1.163 & 1.163 \\
\hline 2010 & 0.143 & 1.002 & 1.192 & 1.198 & 1.199 & 1.199 & 1.199 \\
\hline 2011 & 0.109 & 0.759 & 0.902 & 0.907 & 0.907 & 0.907 & 0.907 \\
\hline 2012 & 0.143 & 0.999 & 1.189 & 1.195 & 1.195 & 1.195 & 1.195 \\
\hline 2013 & 0.082 & 0.576 & 0.686 & 0.689 & 0.689 & 0.689 & 0.689 \\
\hline 2014 & 0.189 & 1.32 & 1.57 & 1.578 & 1.579 & 1.579 & 1.579 \\
\hline 2015 & 0.122 & 0.854 & 1.016 & 1.021 & 1.021 & 1.021 & 1.021 \\
\hline 2016 & 0.071 & 0.494 & 0.588 & 0.591 & 0.591 & 0.591 & 0.591 \\
\hline 2017 & 0.061 & 0.426 & 0.507 & 0.509 & 0.509 & 0.509 & 0.509 \\
\hline 2018 & 0.059 & 0.411 & 0.489 & 0.492 & 0.492 & 0.492 & 0.492 \\
\hline
\end{tabular}

Table 40.10. Whiting7.a Stock Numbers-at-age (start of year) ('1000).
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline & Age 0 & Age 1 & Age 2 & Age 3 & Age 4 & Age 5 & Age 6 \\
\hline 1980 & 635229 & 389102 & 123970 & 20821 & 7385 & 1015 & 1811 \\
\hline 1981 & 322377 & 210420 & 146103 & 39006 & 6731 & 2493 & 987 \\
\hline 1982 & 284961 & 106282 & 76586 & 42548 & 11502 & 2069 & 1107 \\
\hline 1983 & 880745 & 93635 & 37846 & 21124 & 11761 & 3309 & 946 \\
\hline 1984 & 630711 & 289238 & 33220 & 10343 & 5776 & 3346 & 1253 \\
\hline 1985 & 513642 & 205477 & 97366 & 7969 & 2422 & 1403 & 1156 \\
\hline 1986 & 870632 & 165785 & 65062 & 20063 & 1557 & 489 & 534 \\
\hline 1987 & 474781 & 284311 & 56680 & 16221 & 4918 & 396 & 269 \\
\hline 1988 & 488017 & 154654 & 95613 & 13564 & 3787 & 1191 & 167 \\
\hline 1989 & 601264 & 160029 & 54338 & 25509 & 3604 & 1046 & 388 \\
\hline 1990 & 522554 & 194247 & 50982 & 11368 & 5076 & 741 & 305 \\
\hline 1991 & 684435 & 170152 & 65163 & 12125 & 2635 & 1220 & 260 \\
\hline 1992 & 230234 & 222831 & 57027 & 15463 & 2803 & 631 & 367 \\
\hline 1993 & 214824 & 72931 & 62381 & 8655 & 2101 & 390 & 144 \\
\hline 1994 & 183733 & 69063 & 22498 & 12048 & 1566 & 392 & 103 \\
\hline 1995 & 340766 & 58958 & 21047 & 4216 & 2103 & 282 & 92 \\
\hline 1996 & 203244 & 104625 & 12876 & 4367 & 972 & 512 & 94 \\
\hline 1997 & 170729 & 61766 & 21270 & 2453 & 924 & 217 & 140 \\
\hline 1998 & 167298 & 52681 & 13969 & 4600 & 590 & 235 & 94 \\
\hline 1999 & 208166 & 49661 & 9088 & 2189 & 800 & 108 & 63 \\
\hline 2000 & 109119 & 63565 & 10440 & 1802 & 482 & 186 & 41 \\
\hline 2001 & 191982 & 32222 & 10573 & 1566 & 300 & 85 & 41 \\
\hline 2002 & 79962 & 58251 & 6479 & 1988 & 327 & 66 & 29 \\
\hline 2003 & 120096 & 23261 & 8725 & 858 & 292 & 51 & 15 \\
\hline 2004 & 95546 & 37820 & 6065 & 2235 & 245 & 88 & 21 \\
\hline 2005 & 105851 & 26218 & 3767 & 494 & 202 & 23 & 11 \\
\hline 2006 & 155056 & 34042 & 7918 & 1149 & 168 & 72 & 13 \\
\hline 2007 & 103942 & 43533 & 3979 & 781 & 126 & 19 & 10 \\
\hline 2008 & 148998 & 30397 & 6765 & 551 & 120 & 20 & 5 \\
\hline 2009 & 92765 & 44859 & 5789 & 1193 & 108 & 25 & 5 \\
\hline 2010 & 92709 & 27465 & 7599 & 888 & 203 & 19 & 6 \\
\hline 2011 & 151604 & 27332 & 4516 & 1125 & 146 & 35 & 4 \\
\hline 2012 & 79507 & 46281 & 5735 & 893 & 247 & 34 & 10 \\
\hline 2013 & 159742 & 23450 & 7635 & 852 & 147 & 43 & 8 \\
\hline 2014 & 198580 & 50053 & 5903 & 1876 & 233 & 42 & 15 \\
\hline 2015 & 122731 & 55942 & 5990 & 599 & 211 & 28 & 7 \\
\hline 2016 & 71365 & 36959 & 10671 & 1058 & 117 & 44 & 7 \\
\hline 2017 & 108234 & 22626 & 10102 & 2891 & 319 & 37 & 17 \\
\hline 2018 & 174768 & 34652 & 6621 & 2968 & 946 & 110 & 19 \\
\hline 2019 & 118853 & 56071 & 10289 & 1980 & 988 & 332 & 47 \\
\hline
\end{tabular}

Table 40.11. Whiting7.a Stock Summary: weights in tonnes: CatchPred is predicted catch from ASAP. Recruitment at age zero ('1000), Fbar ages (1-3).
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline Year & Lan & Dis & Cat & CatPred & Tsb & Ssb & SsbCv & Recr & RecrCv & Fbar & FbarCv \\
\hline 1980 & 13422 & 3314 & 16737 & 16686.66 & 61420.92 & 32899.73 & 0.325139 & 635228.6 & 0.32582 & 0.378692 & 0.324908 \\
\hline 1981 & 18267 & 3064 & 21331 & 21129.95 & 59643.69 & 43125.71 & 0.245329 & 322376.8 & 0.423325 & 0.44553 & 0.293808 \\
\hline 1982 & 17167 & 801 & 17969 & 17944.82 & 43535.43 & 34607.72 & 0.267068 & 284961.4 & 0.444025 & 0.492468 & 0.329788 \\
\hline 1983 & 10577 & 1829 & 12405 & 12364.61 & 29928.45 & 21969.51 & 0.344828 & 880744.9 & 0.252695 & 0.50039 & 0.388297 \\
\hline 1984 & 11619 & 3380 & 14999 & 14737.24 & 37828.55 & 14978.71 & 0.411684 & 630711 & 0.313234 & 0.613033 & 0.354924 \\
\hline 1985 & 15525 & 2644 & 18169 & 17963.08 & 37041.66 & 22719.89 & 0.29332 & 513641.5 & 0.343469 & 0.744383 & 0.325236 \\
\hline 1986 & 10063 & 2066 & 12129 & 12082.71 & 28818.99 & 18159.02 & 0.321438 & 870631.9 & 0.261661 & 0.579761 & 0.350248 \\
\hline 1987 & 10411 & 3859 & 14270 & 14066.37 & 33168.22 & 16166.41 & 0.336946 & 474780.8 & 0.346118 & 0.615143 & 0.3249 \\
\hline 1988 & 10245 & 1611 & 11856 & 11796.21 & 29995.06 & 20452.89 & 0.278758 & 488017.2 & 0.322468 & 0.521227 & 0.327535 \\
\hline 1989 & 11305 & 2103 & 13408 & 13403.95 & 26698.08 & 16984.32 & 0.306306 & 601264.4 & 0.266655 & 0.731254 & 0.323376 \\
\hline 1990 & 8212 & 2444 & 10656 & 10636.01 & 24487.3 & 12677.07 & 0.327726 & 522554.5 & 0.249696 & 0.620477 & 0.308166 \\
\hline 1991 & 7348 & 2598 & 9946 & 9916.834 & 23051.68 & 13778.43 & 0.261503 & 684435.3 & 0.167116 & 0.622459 & 0.255149 \\
\hline 1992 & 8588 & 4203 & 12791 & 12550.01 & 22568.53 & 11872.65 & 0.20842 & 230234.3 & 0.151858 & 1.008557 & 0.186355 \\
\hline 1993 & 6523 & 2707 & 9230 & 6899.206 & 13307.06 & 9930.362 & 0.151344 & 214824.4 & 0.130265 & 0.800337 & 0.169928 \\
\hline 1994 & 6763 & 1173 & 7936 & 5062.492 & 8920.66 & 5729.961 & 0.161059 & 183732.7 & 0.136829 & 0.826415 & 0.1764 \\
\hline 1995 & 4893 & 2151 & 7044 & 4563.806 & 6915.524 & 4126.811 & 0.163397 & 340766.3 & 0.118301 & 0.810776 & 0.171416 \\
\hline 1996 & 4335 & 3631 & 7966 & 4401.417 & 7272.473 & 2878.211 & 0.185191 & 203244.2 & 0.128363 & 0.891615 & 0.149134 \\
\hline 1997 & 2277 & 1928 & 4205 & 3013.874 & 5369.606 & 3102.811 & 0.150431 & 170729.1 & 0.138898 & 0.771362 & 0.166861 \\
\hline 1998 & 2229 & 1304 & 3533 & 2898.949 & 4268.241 & 2571.897 & 0.158677 & 167298.4 & 0.130187 & 1.076882 & 0.167957 \\
\hline 1999 & 1670 & 1092 & 2762 & 2262.029 & 3105.083 & 1550.688 & 0.195674 & 208166.1 & 0.127571 & 0.853836 & 0.178442 \\
\hline 2000 & 762 & 2118 & 2880 & 2366.311 & 3492.466 & 1509.236 & 0.188354 & 109119 & 0.139041 & 1.118118 & 0.154152 \\
\hline 2001 & 733 & 1012 & 1745 & 1635.71 & 2273.041 & 1328.933 & 0.171596 & 191982.3 & 0.130194 & 0.904033 & 0.184241 \\
\hline 2002 & 747 & 740 & 1487 & 1891.912 & 2871.969 & 1182.682 & 0.192723 & 79961.99 & 0.136065 & 1.236361 & 0.160342 \\
\hline 2003 & 517 & 480 & 996 & 1276.436 & 1848.185 & 1194.555 & 0.170942 & 120095.8 & 0.148765 & 0.61084 & 0.237885 \\
\hline 2004 & 133 & 905 & 1038 & 2078.563 & 2357.977 & 1268.768 & 0.19478 & 95545.88 & 0.124788 & 1.696695 & 0.173514 \\
\hline 2005 & 125 & 272 & 397 & 499.4508 & 1389.403 & 542.5556 & 0.262056 & 105850.6 & 0.138365 & 0.444997 & 0.279148 \\
\hline 2006 & 64 & 1773 & 1837 & 2592.853 & 2205.217 & 1078.421 & 0.21537 & 155056.1 & 0.125793 & 1.516174 & 0.183468 \\
\hline 2007 & 35 & 1512 & 1547 & 1531.212 & 2112.059 & 579.6879 & 0.270007 & 103942.5 & 0.130861 & 1.194777 & 0.170542 \\
\hline
\end{tabular}
\begin{tabular}{lccccccccccc}
\hline Year & Lan & Dis & Cat & CatPred & Tsb & Ssb & SsbCv & Recr & RecrCv & Fbar & FbarCv \\
\hline 2008 & 37 & 1169 & 1206 & 1285.84 & 1669.595 & 724.2601 & 0.207149 & 148998 & 0.124539 & 0.965297 & 0.193505 \\
\hline 2009 & 39 & 1321 & 1360 & 1499.778 & 2220.505 & 735.6692 & 0.212079 & 92764.96 & 0.133154 & 1.097418 & 0.176569 \\
\hline 2010 & 30 & 1154 & 1184 & 1446.459 & 1716.639 & 821.2836 & 0.19131 & 92709.21 & 0.128164 & 1.131069 & 0.190696 \\
\hline 2011 & 31 & 946 & 977 & 985.5709 & 1469.646 & 614.1699 & 0.229025 & 151604 & 0.125392 & 0.855986 & 0.198344 \\
\hline 2012 & 60 & 1339 & 1399 & 1433.651 & 2163.881 & 682.8956 & 0.212126 & 79506.58 & 0.135284 & 1.127338 & 0.173719 \\
\hline 2013 & 33 & 948 & 981 & 1035.359 & 1619.363 & 836.1217 & 0.197555 & 159742 & 0.140032 & 0.650456 & 0.222075 \\
\hline 2014 & 23 & 1951 & 1974 & 2637.666 & 2788.931 & 941.9842 & 0.209781 & 198580.4 & 0.11835 & 1.489569 & 0.179951 \\
\hline 2015 & 28 & 1521 & 1549 & 1727.956 & 2572.353 & 675.9363 & 0.252716 & 122731 & 0.12349 & 0.96351 & 0.188623 \\
\hline 2016 & 15 & 765 & 780 & 816.4469 & 2103.228 & 964.8841 & 0.191329 & 71364.91 & 0.154039 & 0.557581 & 0.224075 \\
\hline 2017 & 36 & 668 & 704 & 728.7112 & 1838.558 & 1146.201 & 0.192279 & 108234.5 & 0.172921 & 0.480626 & 0.263552 \\
\hline 2018 & 46 & 853 & 899 & 921.0501 & 2175.082 & 1076.617 & 0.235776 & 174768 & 0.236359 & 0.464076 & 0.292646 \\
\hline \(2019 *\) & NA & NA & NA & NA & \(N A\) & 1375.539 & \(N A\) & 118853 & \(N A\) & 0.500761 & \(N A\) \\
\hline
\end{tabular}

Table 40.12. Whiting7.a. Input values for short-term forecast. Note that Sel and CWt refer to the landings and DSel and DCWt refer to the discards. Numbers in thousands; Weights in kg.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|}
\hline \multicolumn{11}{|l|}{2019} \\
\hline Age & N & M & Mat & PF & PM & SWt & Sel & CWt & DSel & DCWt \\
\hline 0 & 118853 & 1.078 & 0 & 0 & 0 & 0 & 0 & 0 & 0.063 & 0.024 \\
\hline 1 & 56071 & 0.803 & 0 & 0 & 0 & 0.031 & 0.05 & 0.129 & 0.394 & 0.037 \\
\hline 2 & 10289 & 0.718 & 1 & 0 & 0 & 0.064 & 0.236 & 0.259 & 0.292 & 0.088 \\
\hline 3 & 1980 & 0.608 & 1 & 0 & 0 & 0.147 & 0.398 & 0.311 & 0.132 & 0.179 \\
\hline 4 & 988 & 0.554 & 1 & 0 & 0 & 0.277 & 0.489 & 0.4 & 0.041 & 0.293 \\
\hline 5 & 332 & 0.518 & 1 & 0 & 0 & 0.387 & 0.522 & 0.438 & 0.009 & 0.233 \\
\hline 6 & 47 & 0.518 & 1 & 0 & 0 & 0.505 & 0.524 & 0.615 & 0.007 & 0 \\
\hline \multicolumn{11}{|l|}{2020} \\
\hline Age & N & M & Mat & PF & PM & SWt & Sel & CWt & DSel & DCWt \\
\hline 0 & 118853 & 1.078 & 0 & 0 & 0 & 0 & 0 & 0 & 0.063 & 0.024 \\
\hline 1 & 37954 & 0.803 & 0 & 0 & 0 & 0.031 & 0.05 & 0.129 & 0.394 & 0.037 \\
\hline 2 & 16117 & 0.718 & 1 & 0 & 0 & 0.064 & 0.236 & 0.259 & 0.292 & 0.088 \\
\hline 3 & 2960 & 0.608 & 1 & 0 & 0 & 0.147 & 0.398 & 0.311 & 0.132 & 0.179 \\
\hline 4 & 634 & 0.554 & 1 & 0 & 0 & 0.277 & 0.489 & 0.4 & 0.041 & 0.293 \\
\hline 5 & 334 & 0.518 & 1 & 0 & 0 & 0.387 & 0.522 & 0.438 & 0.009 & 0.233 \\
\hline 6 & 133 & 0.518 & 1 & 0 & 0 & 0.505 & 0.524 & 0.615 & 0.007 & 0 \\
\hline \multicolumn{11}{|l|}{2021} \\
\hline Age & \(N\) & M & Mat & PF & PM & SWt & Sel & CWt & DSel & DCWt \\
\hline 0 & 118853 & 1.078 & 0 & 0 & 0 & 0 & 0 & 0 & 0.063 & 0.024 \\
\hline 1 & 37954 & 0.803 & 0 & 0 & 0 & 0.031 & 0.05 & 0.129 & 0.394 & 0.037 \\
\hline 2 & 10910 & 0.718 & 1 & 0 & 0 & 0.064 & 0.236 & 0.259 & 0.292 & 0.088 \\
\hline 3 & 4636 & 0.608 & 1 & 0 & 0 & 0.147 & 0.398 & 0.311 & 0.132 & 0.179 \\
\hline 4 & 948 & 0.554 & 1 & 0 & 0 & 0.277 & 0.489 & 0.4 & 0.041 & 0.293 \\
\hline 5 & 214 & 0.518 & 1 & 0 & 0 & 0.387 & 0.522 & 0.438 & 0.009 & 0.233 \\
\hline 6 & 164 & 0.518 & 1 & 0 & 0 & 0.505 & 0.524 & 0.615 & 0.007 & 0 \\
\hline
\end{tabular}

Table 40.13. Whiting7.a. Single-option output of the short-term forecast. ( \(F=\) mean \(F\) 2016-2018). Numbers in thousands, weights in tonnes.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|}
\hline \multicolumn{11}{|l|}{2019} \\
\hline Age & F & CatchNos & Yield & DF & DCatchNos & DYield & StockNos & Biomass & SSNos & SSB \\
\hline 0 & 0 & 0 & 0 & 0.063 & 4500 & 107 & 118853 & 0 & 0 & 0 \\
\hline 1 & 0.05 & 1 & 0 & 0.394 & 14220 & 531 & 56071 & 1740 & 0 & 0 \\
\hline 2 & 0.236 & 9 & 2 & 0.292 & 3097 & 273 & 10289 & 659 & 10289 & 659 \\
\hline 3 & 0.398 & 71 & 22 & 0.132 & 556 & 100 & 1980 & 291 & 1980 & 291 \\
\hline 4 & 0.489 & 153 & 61 & 0.041 & 167 & 49 & 988 & 274 & 988 & 274 \\
\hline 5 & 0.522 & 97 & 42 & 0.009 & 13 & 3 & 332 & 128 & 332 & 128 \\
\hline 6 & 0.524 & 16 & 10 & 0.007 & 0 & 0 & 47 & 24 & 47 & 24 \\
\hline Total & 0.228 & 347 & 137 & 0.273 & 22553 & 1063 & 188560 & 3116 & 13636 & 1376 \\
\hline \multicolumn{11}{|l|}{2020} \\
\hline Age & F & CatchNos & Yield & DF & DCatchNos & DYield & StockNos & Biomass & SSNos & SSB \\
\hline 0 & 0 & 0 & 0 & 0.063 & 4500 & 107 & 118853 & 0 & 0 & 0 \\
\hline 1 & 0.05 & 0 & 0 & 0.394 & 9625 & 359 & 37954 & 1178 & 0 & 0 \\
\hline 2 & 0.236 & 14 & 4 & 0.292 & 4851 & 427 & 16117 & 1031 & 16117 & 1031 \\
\hline 3 & 0.398 & 107 & 33 & 0.132 & 831 & 149 & 2960 & 435 & 2960 & 435 \\
\hline 4 & 0.489 & 98 & 39 & 0.041 & 107 & 31 & 634 & 176 & 634 & 176 \\
\hline 5 & 0.522 & 97 & 43 & 0.009 & 13 & 3 & 334 & 129 & 334 & 129 \\
\hline 6 & 0.524 & 44 & 27 & 0.007 & 0 & 0 & 133 & 67 & 133 & 67 \\
\hline Total & 0.228 & 360 & 146 & 0.273 & 19927 & 1076 & 176985 & 3016 & 20178 & 1838 \\
\hline \multicolumn{11}{|l|}{2021} \\
\hline Age & F & CatchNos & Yield & DF & DCatchNos & DYield & StockNos & Biomass & SSNos & SSB \\
\hline 0 & 0 & 0 & 0 & 0.063 & 4500 & 107 & 118853 & 0 & 0 & 0 \\
\hline 1 & 0.05 & 0 & 0 & 0.394 & 9625 & 359 & 37954 & 1178 & 0 & 0 \\
\hline 2 & 0.236 & 9 & 2 & 0.292 & 3283 & 289 & 10910 & 698 & 10910 & 698 \\
\hline 3 & 0.398 & 167 & 52 & 0.132 & 1302 & 233 & 4636 & 682 & 4636 & 682 \\
\hline 4 & 0.489 & 147 & 59 & 0.041 & 160 & 47 & 948 & 262 & 948 & 262 \\
\hline 5 & 0.522 & 62 & 27 & 0.009 & 8 & 2 & 214 & 83 & 214 & 83 \\
\hline 6 & 0.524 & 54 & 33 & 0.007 & 0 & 0 & 164 & 83 & 164 & 83 \\
\hline Total & 0.228 & 439 & 173 & 0.273 & 18878 & 1037 & 173679 & 2986 & 16872 & 1808 \\
\hline
\end{tabular}

Table 40.14. Whiting7.a. Management options table. Weights in tonnes.
\begin{tabular}{|l|l|l|l|l|l|l|l|l|l|l|}
\hline Fmult & Catch20 & Land20 & Dis20 & Basis & FCatch20 & FLand20 & FDis20 & SSB21 & dSSB & dTac \\
\hline 0 & 0 & 0 & 0 & & 0 & NA & NA & 2971 & \(61.56 \%\) & \(-100 \%\) \\
\hline 0.1 & 146 & 18 & 128 & & 0.05008 & 0.00206 & 0.04801 & 2827 & \(53.73 \%\) & \(-97.52 \%\) \\
\hline 0.2 & 286 & 35 & 251 & & 0.10015 & 0.00412 & 0.09603 & 2690 & \(46.28 \%\) & \(-95.19 \%\) \\
\hline 0.3 & 420 & 51 & 369 & & 0.15023 & 0.00619 & 0.14404 & 2559 & \(39.15 \%\) & \(-92.99 \%\) \\
\hline 0.4 & 549 & 67 & 482 & & 0.2003 & 0.00825 & 0.19206 & 2435 & \(32.41 \%\) & \(-90.78 \%\) \\
\hline 0.5 & 672 & 81 & 591 & & 0.25038 & 0.01031 & 0.24007 & 2317 & \(25.99 \%\) & \(-88.86 \%\) \\
\hline 0.6 & 791 & 95 & 696 & & 0.30046 & 0.01237 & 0.28809 & 2205 & \(19.90 \%\) & \(-86.93 \%\) \\
\hline 0.7 & 905 & 109 & 796 & & 0.35053 & 0.01443 & 0.3361 & 2098 & \(14.08 \%\) & \(-85.01 \%\) \\
\hline 0.8 & 1015 & 122 & 893 & & 0.40061 & 0.0165 & 0.38411 & 1996 & \(8.54 \%\) & \(-83.22 \%\) \\
\hline 0.9 & 1120 & 134 & 986 & & 0.45068 & 0.01856 & 0.43213 & 1900 & \(3.32 \%\) & \(-81.57 \%\) \\
\hline 1 & 1222 & 146 & 1076 & & 0.50076 & 0.02062 & 0.48014 & 1808 & \(-1.69 \%\) & \(-79.92 \%\) \\
\hline 1.1 & 1319 & 157 & 1162 & & 0.55084 & 0.02268 & 0.52816 & 1720 & \(-6.47 \%\) & \(-78.40 \%\) \\
\hline 1.2 & 1413 & 167 & 1246 & & 0.60091 & 0.02474 & 0.57617 & 1637 & \(-10.98 \%\) & \(-77.03 \%\) \\
\hline 1.3 & 1503 & 177 & 1326 & & 0.65099 & 0.0268 & 0.62418 & 1558 & \(-15.28 \%\) & \(-75.65 \%\) \\
\hline 1.4 & 1590 & 187 & 1403 & & 0.70107 & 0.02887 & 0.6722 & 1483 & \(-19.36 \%\) & \(-74.28 \%\) \\
\hline 1.5 & 1674 & 196 & 1477 & & 0.75114 & 0.03093 & 0.72021 & 1411 & \(-23.27 \%\) & \(-73 \%\) \\
\hline 1.6 & 1754 & 205 & 1549 & & 0.80122 & 0.03299 & 0.76823 & 1343 & \(-26.97 \%\) & \(-71.80 \%\) \\
\hline 1.7 & 1832 & 214 & 1618 & & 0.85129 & 0.03505 & 0.81624 & 1278 & \(-30.51 \%\) & \(-70.56 \%\) \\
\hline 1.8 & 1907 & 222 & 1685 & & 0.90137 & 0.03711 & 0.86426 & 1216 & \(-33.88 \%\) & \(-69.46 \%\) \\
\hline 1.9 & 1979 & 229 & 1750 & & 0.95145 & 0.03918 & 0.91227 & 1158 & \(-37.03 \%\) & \(-68.50 \%\) \\
\hline 2 & 2048 & 237 & 1812 & & 1.00152 & 0.04124 & 0.96028 & 1102 & \(-40.08 \%\) & \(-67.40 \%\) \\
\hline 2.1 & 2115 & 244 & 1872 & & 1.0516 & 0.0433 & 1.0083 & 1049 & \(-42.96 \%\) & \(-66 \%\) \\
\hline 2.2 & 2180 & 250 & 1930 & & 1.10167 & 0.04536 & 1.05631 & 998 & \(-45.73 \%\) & \(-65.61 \%\) \\
\hline 2.3 & 2243 & 257 & 1986 & & 1.15175 & 0.04742 & 1.10433 & 950 & \(-48.34 \%\) & \(-64.65 \%\) \\
\hline 2.4 & 2303 & 263 & 2040 & & 1.20183 & 0.04949 & 1.15234 & 904 & \(-50.84 \%\) & \(-63.82 \%\) \\
\hline 2.5 & 2361 & 269 & 2093 & & 1.2519 & 0.05155 & 1.20036 & 861 & \(-53.18 \%\) & \(-63.00 \%\) \\
\hline & & & & & & & & & \\
\hline
\end{tabular}


Figure 40.1. Whiting 7.a. Working group estimates of International landings and discards between 1980-2018.


Figure 40.2. 7.a Whiting discard length-frequency by national fleets in 2018 for the OTB_CRU métier. Note due to low levels of retained catch, and hence low sampling, these data are not presented.


Figure 40.3. Whiting 7.a Proportion of discards by age (left) and year (right).


Figure 40.4. Whiting 7.a Smoothed Stock Weights (Three year running average).


Figure 40.5. Whiting 7.a. Log Standardized indices of tuning fleets by cohort.


Figure 40.6. Whiting 7.a. Selectivity-at-age in the Catch.


Figure 40.7. Whiting 7.a. Observed and Predicted Catches.


Figure 40.8. Whiting 7.a. Observed and Predicted index cpue.





Figure 40.9. Whiting 7.a. Retrospective analysis of the final ASAP run with Mohn's Rho calculation.


Figure 40.10. Whiting 7.a. Stock Summary Plot. The thick black line represents the ASAP assessment standard deviations from ASAP are shaded grey. The forecast/assumed values are given by open circles. The thick black line in the catch plot represents the predicted catch from ASAP.


Figure 40.11. Whiting 7.a. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.

\section*{39 Whiting (Merlangius merlangus) in divisions 7.b-c and 7.e-k (southern Celtic Seas and eastern English Channel)}

\section*{Type of assessment in 2019}

Full analytical assessment (XSA) and short-term forecast tuned with a single combined survey index according to the stock annex. Since WGCSE 2015, national discard data have been available through InterCatch for countries with significant landings for this stock. Biological reference points proposed by WKMSYREF4 (ICES, 2016a) are included also.

\section*{ICES advice applicable to 2019}

ICES advises that when the MSY approach is applied, catches in 2019 should be no more than 15841 tonnes.
Management should be implemented at the stock level.
http://ices.dk/sites/pub/Publication\%20Reports/Advice/2018/2018/whg.27.7b-ce-k.pdf

\section*{ICES advice applicable to 2018}

ICES advises that when the MSY approach is applied, catches in 2018 should be no more than 19429 tonnes.
http://ices.dk/sites/pub/Publication\%20Reports/Advice/2017/2017/whg.27.7b-ce-k.pdf

\subsection*{39.1 General}

\section*{Stock description and management units}

The TAC for whiting is set for divisions 7.b, 7.c, 7.d, 7.e, 7.f, 7.g, 7.h, 7.j and 7.k. The assessment area does not correspond to the TAC area. Since the 2014 Benchmark (WKCELT), Whiting in 7.b,c are now assessed as part of 7.bc, e-k, while whiting in 7.d remain part of the WGNSSK assessment of the North Sea stock. Any management measures implemented for this stock should be consistent with the assessment area.


The TAC for whiting 7.bc, e-k decreased from 22213 t (2018) to 19184 (2019). ICES official landings for whiting 7.bc, e-k in 2019 are 9019 t and estimated catch of 10268 t . Thus, the current TAC for whiting catches in the \(7 . \mathrm{bk}\) stock area is not restrictive in the \(7 . \mathrm{bc}, \mathrm{e}-\mathrm{k}\) assessment area.

TAC in 2018/120
\begin{tabular}{|c|c|c|c|}
\hline Species: & \begin{tabular}{l}
Whiting \\
Merlangius merlangus
\end{tabular} & Zone: & \(7 \mathrm{~b}, 7 \mathrm{c}, 7 \mathrm{~d}, 7 \mathrm{e}, 7 \mathrm{f}, 7 \mathrm{~g}, 7 \mathrm{~h}, 7 \mathrm{j}\) and 7 k (WHG/7X7A-C) \\
\hline Belgium & 217 & & \\
\hline France & 13328 & & \\
\hline Ireland & 6176 & & \\
\hline The Netherlands & - 108 & & \\
\hline United Kingdom & 1 2384 & & \\
\hline Union & 22213 & & \\
\hline TAC & 22213 & & \begin{tabular}{l}
Analytical TAC \\
Article 7(2) of this Regulation applies Article 12(1) of this Regulation applies
\end{tabular} \\
\hline
\end{tabular}
\begin{tabular}{|c|c|c|c|}
\hline Species: & Whiting Merlangius merlangus & Zone: & \(7 \mathrm{~b}, 7 \mathrm{c}, 7 \mathrm{~d}, 7 \mathrm{e}, 7 \mathrm{f}, 7 \mathrm{~g}, 7 \mathrm{~h}, 7 \mathrm{j}\) and 7 k (WHG/7X7A-C) \\
\hline Belgium & 187 & & \\
\hline France & 11510 & & \\
\hline Ireland & 5334 & & \\
\hline The Netherlands & 94 & & \\
\hline United Kingdom & 2059 & & \\
\hline Union & 19184 & & \\
\hline TAC & 19184 & & \begin{tabular}{l}
Analytical TAC \\
Article 7(2) of this Regulation applies \\
Article 13(1) of this Regulation applies
\end{tabular} \\
\hline
\end{tabular}

\section*{Landings obligation}

Since 2017, the landings obligation (LO) has applied to this stock in accordance with Delegated Regulation (EC, 2016) superseded by (EU) 2018/2034․ This implies that all catches of whiting in the Celtic Sea and Western Channel by those vessels must be landed. However, a \(6 \%\) de minimus applies to bottom trawls using a mesh size of \(\geq 80 \mathrm{~mm}\), as well as pelagic trawls and beam trawls using \(80-119 \mathrm{~mm}\) mesh. There are also three specific technical measures in operation for vessels using bottom trawls or seines in the Celtic Sea Protection Zone.

A significant proportion of unwanted catch is above the Minimum Conservation Reference Size \((M C R S=27 \mathrm{~cm})\) in whiting, although discards are assessed by ICES to have reduced in 2018 to \(15 \%\) from \(32-26 \%\) for the previous three years. Whiting is also the least limiting stock for most fleets in a mixed fishery context for the Celtic Sea, where cod is generally considered the likely choke species. In this context it is difficult to accurately predict the impact of the LO on Celtic Sea whiting.

\footnotetext{
\({ }^{1}\) https://www.agriculture.gov.ie/media/migration/seafood/fisheriesmanagementnotices/2019/CommissionDelRegEU2034NWW241218.pdf
}

\subsection*{39.2 The fishery in 2018}

ICES officially reported landings for divisions \(7 . \mathrm{b}, \mathrm{c}, \mathrm{e}-\mathrm{k}\) and landings as used by the Working Group are given in Table 1.

Catch for 7.b, c , \(\mathrm{e}-\mathrm{k}\) in addition to landings for 7.d (excluding discards) is also presented as a guide figure for comparison to the 7.b-k TAC.

The 7.bc, e-k whiting stock is primarily targeted by otter trawlers and to a lesser extent Scottish seines and beam trawls. An overview of landings by fleet is given in Table 2 and more generally effort trends in fleets catching whiting in the Celtic Sea is provided by STECF (STECF, 2018).

The spatial distributions of landings by Irish fleets 1995-2016 are given in Figure 1. Irish catches are primarily from within 7.g particularly within 32E2 and 31E3. Landings also emanate, to a lesser extent from 7.j. In previous years, French landings have exhibited similar spatial and temporal focus around 31E3. The majority of UK landings are from otter trawlers in 7.e, and concentrated within rectangles 29E5 and 29E6.

\subsection*{39.3 Data}

\section*{Landings}

National landings and numbers-at-age data were aggregated in InterCatch for the Area 7.bc, ek following methodology described in the stock annex.

The allocation schemes below were used:
Discard raising scheme
\begin{tabular}{|c|c|c|c|}
\hline STRATA & UNSAMPLED & & SAMPLED \\
\hline 1 & GNS_UK & -> & GNS_UK \\
\hline 2 & GNS_FRA & -> & GNS_FRA \\
\hline 3 & TBB_BEL\&UK & -> & TBB_UK \\
\hline 4 & OTT_DEF_70_99_FRA & -> & OTT_DEF_70_99_FRA \\
\hline 5 & OTT_100_119_FRA & -> & OTT_100_119_FRA \\
\hline 6 & GTR_DEF_FRA & -> & GTR_DEF_FRA[1] \\
\hline 7 & GTR_DEF_UK & -> & GTR_DEF_UK \\
\hline 8 & SSC_AllCountries_7e & -> & SSC_IRL_7g \\
\hline 9 & OTM_PTM_7ej & -> & OTB_DEF_AllCountries_7egh \\
\hline 10 & MIS_UK & -> & MIS_UK \\
\hline 11 & MIS_IRL & -> & GNS_IRL \\
\hline 12 & MIS_FRA & -> & GNS_FRA \\
\hline 13 & LLS_UK & -> & GNS_UK \\
\hline 14 & OTB_DEF_70_99_IRL_FRA & -> & OTB_DEF_70_99_FRA_UK \\
\hline 15 & OTB_DEF_>=100_AllCountriesAreas & -> & OTB_DEF_ \\
\hline 16 & OTB_CRU_16_100_AllCountriesAreas & -> & OTB_DEF_70_100_FRA_UK \\
\hline 17 & OTB_CRU_>=100_IRL_UK & -> & OTB_DEF_100_FRA_IRL \\
\hline
\end{tabular}

Sample allocation scheme
\begin{tabular}{|c|c|c|c|}
\hline Strata & Unsampled & \multicolumn{2}{|l|}{Sampled} \\
\hline 1 & GNS_UK & -> & GNS_UK \\
\hline 2 & GNS_FRA & -> & GNS_FRA \\
\hline 3 & TBB_BEL\&UK & -> & TBB_UK \\
\hline 4 & OTT_DEF_70_99_FRA & -> & OTT_DEF_70_99_FRA \\
\hline 5 & OTT_100_119_FRA & -> & OTT_100_119_FRA \\
\hline 6 & GTR_DEF_FRA & -> & GTR_DEF_FRA[1] \\
\hline 7 & GTR_DEF_UK & -> & GTR_DEF_UK \\
\hline 8 & SSC_AllCountries_7e & -> & SSC_IRL_7g \\
\hline 9 & OTM_PTM_7ej & -> & OTB_DEF_AllCountries_7egh \\
\hline 10 & MIS_UK & -> & MIS_UK \\
\hline 11 & MIS_IRL & -> & GNS_IRL \\
\hline 12 & MIS_FRA & -> & GNS_FRA \\
\hline 13 & LLS_UK & -> & GNS_UK \\
\hline 14 & OTB_DEF_70_99_IRL_FRA & -> & OTB_DEF_70_99_FRA_UK \\
\hline 15 & OTB_DEF_>=100_AllCountriesAreas & -> & OTB_DEF_ \\
\hline 16 & OTB_CRU_16_100_AllCountriesAreas & -> & OTB_DEF_70_100_FRA_UK \\
\hline 17 & OTB_CRU_>=100_IRL_UK & -> & OTB_DEF_ \\
\hline
\end{tabular}

NB: Everything has been raised by CATON.

\section*{Age sampling allocation scheme}

The length compositions available in InterCatch for 2018 from the main gears are presented in Table 3 and Figure 2. The landings and discard length distributions are similar for the all otter trawl fleets (OTB), but TBB tend to have discarded slightly larger fish. It is important to note that discards for the French OTB fleet are significantly reduced in 2018 relative to previous years.

The international catch and landings numbers-at-age are given in Table 4 and Figure 3. It is possible to track the very strong 1999 and 2013 year classes, but the strong 2009 recruitment is only apparent at some older ages. The age distribution had remained similar over time with the exception of periods where strong year classes pass through older ages. In 2018 however, there appears to be a significant shortfall in 1+ age classes in both the survey sampling (Figure 7) and commercial data from France in particular (Figure 2).

Age group-0 was included in the assessment data to allow inclusion of -group indices in the XSA, although landings at this age are not recorded in most years. Mean weights-at-age in the catch and stock (Tables 5 and 6) were derived as per methodology described in the stock annex. The stock weights are shown in Figure 4. There is some variability of stock weights particularly at older ages. Mean weight-at-age appears to have declined during the period of recent high fishing effort and landings between 2005-2008.

\section*{Discards}

A time-series of discard data for Ireland and France was made available at WKCELT 2014 and is now included in the assessment. Procedures for raising discards to international landings are
described above and in the stock annex. However, as more accurate national data become available through InterCatch, these have been included in the assessment as an improvement over simply raising Irish and French OTB discards to the international landings to produce a catch time-series.

A summary of discarding rates-at-age for 2018 as available in InterCatch is presented in Table 7. Discarded whiting length distributions from 2017 for the main fleets is presented in Figure 2. The available data suggest that discarding occurs well above the 27 cm minimum conservation reference size (MCRS) with fish occasionally being discarded above 40 cm in some fleets. Annual proportions-at-age of discard numbers in the catch, and also catch numbers in the predicted stock from the XSA assessment are given in Figure 3. Data show a recent upward trend in discarding of all ages in the catch and stock.

Figure 5 presents the proportion of 1-3 year olds in the discards vs stock, discard vs catch and catch vs stock respectively. The data suggest that the ratio of 1:2 year old fish in the discards vs catch have remained relatively constant to each other. However, the proportion of three-year old fish being taken in the catch and discards has increased in 2018 relative to two-year old fish in the stock meaning either F for that age group probably has not gone up, but the estimates of that cohort in the stock has proportionately come down.

\section*{Biological}

Mean stock and catch weights-at-age data were calculated following the methodology described in the stock annex. Natural mortality is based on Lorenzen's model and thus a power function of catch weights-at-age. Maturity is knife-edge at-age 2.

The proportions of F and M before spawning were both set to zero to reflect the SSB calculation date of January 1st.

\section*{Surveys}

The combined Q4 IBTS survey index for the Irish (IGFS) and French (EVHOE) time-series for ages \(0-5\) is given in Table 8. Further details for combining the survey series is given in the stock annex.

The internal consistency of the survey tuning fleet was examined using pairwise scatterplots of log numbers-at-age (Figure 6), bearing in mind that the correlations may be impacted by changes in fishing mortality. Other than 0 vs 1-grp fish, the index is reasonably consistent for older ages (ages 1-5).

Cohort and year effects were examined with mean log standardized plots of indices by cohort Figure 6 and year (Figure 7). The index is quite noisy and shows a strong year affect for 2018 where all ages above 0-group appear to have dropped significantly.

Examination of the raw length frequency data going into the survey index (see below) from both the IGFS (left panel) and EVHOE (right panel) surveys highlight virtually a complete absence of length data for 2018 above about 20 cm which approximates to an 0 -group fish cut-off point.


This is also the case for French commercial OTB data (Figure 2) and these older age classes appear only to be represented in significant numbers in the Irish commercial data, particularly west of Ireland.

\section*{Commercial Ipue}

Commercial lpue, from 2000 to 2013, were evaluated at WKCELT 2014 and have been omitted from the assessment due to catchability trends.

\subsection*{39.4 Historical stock development}

An XSA assessment was carried out for this stock applying the same settings as last year, using a truncated time-series 1999-2018 of combined landings and discards data. The settings previously used were applied again this year and are detailed within the stock annex.

\section*{Data screening and Final update assessment}

The general methodology is outlined in Section 2. Exploratory analysis was carried out using FLR under R version 3.1.1. The packages FLCore 2.5, and FLXSA 2.5 and FLEDA 2.5 were used.
\begin{tabular}{llr}
\hline & & \\
\hline Catch date range: & Years & \(1999-2017\) \\
\cline { 3 - 3 } Fbar Age Range: & Ages & \(-7+\) \\
\hline Assessment Method: & Yrs & \(2-5\) \\
\hline Survey Tuning-series: & Ages & XSA \\
\hline IGFS-EVHOE & & \(2003-2015\) \\
\hline Time taper: & & -5 \\
\hline F plateau age: & & No \\
\hline & & 5 \\
\hline
\end{tabular}

The full XSA diagnostics are given in Table 9. Overall the estimates are reasonably consistent for ages \(1+\) given that whiting are prone to year effects in survey catches.

The log-catchability residuals from the XSA fit are plotted for the tuning-series in Figure 8. The residual patterns for the survey index shows some significant recent trends. There are strong positive residuals for 0-group and 1-group fish in 2017 changing to negative residuals for 1 and two-year old fish in 2018.

The retrospective pattern is shown in Figure 9. A retrospective bias in F appeared to be developing in this assessment with F being revised down. As with 2017, this year the WG scaled F to F2018 to address this retrospective trend.

In the three most recent years however, there has been a general upward revision in F and downward revision in SSB. In this year's assessment, a significant revision in recent recruitment due to low 1-3 year fish in the catch and survey data are likely to have driven the strong downward revision in SSB.

Estimates of fishing mortality and stock numbers from the final XSA are given in Table 10 and Table 11. These are summarized in Figure 10. The assessment this year reveals a further increase in fishing mortality for older fish and recruitment in 2013 is estimated to be the third highest in the time-series (Figure 11).

\section*{Comparison with previous assessments}

The current assessment shows a strong retrospective revision in terms of recruitment in particular from the 2017 assessment. SSB is also assessed to now be below Blim for the first time in the time-series.


Mohn's Rho was calculated using the recent five years of data, up to and including 2018, to estimate the bias. The estimates gave an \(\mathrm{F}_{\mathrm{bar}}\) value of 0.009 and SSB of 0.061 .

\subsection*{39.5 State of the stock}

Trends in landings, \(\mathrm{F}(2-5)\), SSB , and recruitment are presented in Table 12. For the current timeseries SSB displays a peak biomass in 2012 following the strong recruitment of the 2009 year class and again in 2015 following the 2013 recruitment.

Fishing mortality ( \(\mathrm{F}_{\mathrm{bar}}\) ) has increased since 2012 and is now assessed to be above \(\mathrm{F}_{\text {msy }}\). SSB is now estimated to be below \(\mathrm{B}_{\mathrm{lim}}\), below the precautionary limits for this stock.

The 2013 cohort, estimated to be the third highest in the time-series, now seems to have passed through the fishery without being evident in 2018 in substantial numbers resulting in both recruitment and SSB being revised down in the assessment and \(\mathrm{F}_{\mathrm{bar}}\) revised upwards.

\subsection*{39.6 Short-term projections}

The short-term projection settings were as described in the stock annex with the following exceptions. The GM period was 1999-2017 (full time-series minus the last year).

Table 13 gives the management option table. Given the status of SSB the ICES Advice Rule was
 implies catches of 6481 t and landings of 3885 t .

The input values for the catch forecast (using FLR 2.5) are given in Table 14. The F-at-age values used were calculated as the mean of the XSA values from 2016-2018, scaled to the most recent year. Historically F has been used unscaled, but as mentioned in the Annex, it was suggested in the benchmark that other options might be considered depending on consistent patterns in the retrospective analysis. Catch and stock weights-at-age were also the mean of the period 20162018. Stock numbers-at-age in 2018 for ages and older were obtained from the XSA. SSB values are calculated for 1 January.

The estimated contributions of recent recruited year classes to the landings and SSB predictions are given in Figure 12. Yield is still heavily reliant on the XSA estimate 2018 year class, which is estimated at \(35 \%\). The assumptions of GM1999-2017 recruitment for 2019 are predicted to contribute ca. \(57 \%\) to the SSB in 2021.

\subsection*{39.7 MSY evaluations and Biological reference points}

ICES carried out an evaluation of MSY and PA reference points for this stock at WKMSYREF4 (ICES, 2016a). The results are summarised below:

\section*{Reference points}
\begin{tabular}{|c|c|c|c|c|}
\hline Framework & Reference point & Value & Technical basis & Source \\
\hline \multirow[t]{2}{*}{MSY approach} & MSY \(\mathrm{B}_{\text {trigger }}\) & 35000 t & \(\mathrm{B}_{\mathrm{pa}}\) & ICES, 2016b \\
\hline & \(\mathrm{F}_{\text {MSY }}\) & . 524 & Segmented regression with \(\mathrm{B}_{\lim }\) as the breakpoint Range \(=.32-.67\) & ICES, 2016b \\
\hline \multirow[t]{4}{*}{Precautionary approach} & \(\mathrm{Blim}_{\text {l }}\) & 25000 t & \(\mathrm{B}_{\text {loss, }}\), the lowest observed spawning-stock biomass. & 2016a \\
\hline & \(\mathrm{B}_{\mathrm{pa}}\) & 35000 t & \(\mathrm{Blim}_{\lim } 1.4\) & \[
\begin{aligned}
& \text { ICES, } \\
& \text { 2016a }
\end{aligned}
\] \\
\hline & \(\mathrm{F}_{\text {lim }}\) & 1.120 & Based on segmented regression simulation of recruitment with \(\mathrm{B}_{\text {lim }}\) as the breakpoint & 2016a \\
\hline & \(\mathrm{F}_{\mathrm{pa}}\) & . 800 & \(\mathrm{F}_{\text {lim }} / 1.4\) & 2016a \\
\hline \multirow[t]{2}{*}{Management plan} & SSB \({ }_{\text {MGT }}\) & Undefined & & \\
\hline & \(\mathrm{F}_{\text {MGT }}\) & Undefined & & \\
\hline
\end{tabular}

\subsection*{39.8 Management plans}

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to demersal stocks including whiting in ICES divisions 7.b-ce-k.

\subsection*{39.9 Uncertainties and bias in assessment and forecast}

\section*{Sampling}

Sampling levels of the landed catch for recent years are considered to be sufficient to support current assessment approaches. There has been SOP differences in some recent years particularly that have led to a disparity between the reported catch in tons (landings and discards) going into the assessment and the comparable \(\sum\) (CNAA \(\times\) MWAA) coming out of the assessment. While the overall SOP checks have invariably been \(\leq 3 \%\), any difference in the catches going into the assessment vs those coming out will cause concern. Rather than correct the national data provided therefore a SOP correction is now done within FLR once the initial data QC is complete to ensure corrections are minor and not masking a potential error/bias. Data submitted to InterCatch for 2018 for French discards was unusually low however, with the proportion of discards in the French OTB fleet catch dropping from \(26 \%\) in 2017 to \(7 \%\) in 2018. There is some evidence of very low numbers of 1-4 year olds in the survey tuning index also however.

\section*{Ageing}

Cohort tracking in the landings-at-age matrix appears fairly consistent up to age 6. Tracking deteriorates at older ages.

\section*{Discards}

Discarding is a major feature of most fisheries catching whiting in the Celtic Sea. Sampling coverage of discarding has improved over time particularly since 2004. Attempts to reconstruct a time-series for the main Irish and French fleets failed to extend further back than 1999. No discard data were available for France prior to 2004 and had to be constructed as proportion-at-age for the recent years where data were available. Sampling levels for either country also did not allow for quarterly age-based reconstruction of the discards so a length-based ogive from Ireland had to be used to reconstruct the data for both countries. Discard estimates for the UK were not available at the benchmark, but are available now through InterCatch and have been included in the assessment.

\section*{Selectivity}

Square-mesh panels were introduced in the second half of 2012 to reduce catches and discards of smaller whiting and haddock. The current assessment does not show an obvious reduction in F-at-age since the introduction of this TCM (see Figures 5 and 10).

\section*{Surveys}

The surveys for whiting are prone to year effects. However, cohort tracking for the \(1+\) fish is generally quite consistent for the combined tuning index. As mentioned there is a strong year affect in 2018 for \(1+\) fish which seems somewhat reflected in the commercial OTB data for France at least.

\section*{Misreporting}

The level of misreporting of this stock is not known and underreporting has previously been considered unlikely to have been a significant source of unaccounted mortality of whiting in the assessment because the TAC has been in excess of recent landings.

\subsection*{39.10 Recommendation for next benchmark}

This stock is part of a benchmark process, due to complete in Q1 2020.

\subsection*{39.11 Management considerations}

Catches and SSB in 7.b, c, e-k whiting fluctuate considerably depending on year-class strength. The 2008 and 2009 year classes were above average with 2013 being third highest in the timeseries. These contributed to catches and SSB in the short term but the upturn in catches and SSB was short-lived as recruitment is episodic, and F is now above FMSY and SSB below \(B_{\text {trigger }}\).

Discarding in this stock for different fleets is substantial and highly variable depending on gear and year-class strength. High levels of discarding for a species like whiting reduce the longer term yields one might expect so efforts to improve selection and reduce discards in the mixed fishery should be encouraged. ICES notes the introduction of square mesh panels in all trawl fisheries operating in ICES divisions 7.fg. It is important that these measures are fully implemented and their effectiveness in reducing discards and the impact on commercial catches is monitored and evaluated. Further gear modifications to increase the likelihood of small whiting passing through the gear, such as introduction of larger minimum mesh sizes, separator panels, or grids may be needed.

Ireland has the only directed fishery for whiting, which is part of mixed fishery throughout the Celtic Sea, as well as bycatch within Nephrops fisheries. Discard rates are high as a consequence
of the low market value of the species, particularly at smaller sizes. High-grading above the MCRS to some extent is also prevalent in most fisheries.

From the 1 February to the 31 March, fishing activity has been prohibited within ICES rectangles: 30E4, 31E4, 32E3 (excluding within six nautical miles from the baseline) annually since 2005 to protect the cod stock.
There have been major changes in fleet dynamics over the period of the assessment. Effort in the French gadoid fleet has been declining since 1999, but the effort has fluctuated in recent years due to the way the effort series is derived. Irish otter-trawl effort in \(7 . \mathrm{b}-\mathrm{k}\) has been declined slightly over the time-series.

The full impact of the Landings Obligation is complex and unknown as yet and will depend on whether there is a measurable impact on discarding behaviour or whether variable practices continue and simply data becomes more reliable (for a summary of issues see http://www.discardless.eu/media/results/Celtic Sea Year2.pdf).

\subsection*{39.12 References}

EC. 2015. Commission Delegated Regulation (EU) 2015/2438 of 12 October 2015 establishing a discard plan for certain demersal fisheries in north-western waters.

EC. 2016. Commission Delegated Regulation (EU) 2016/2375 of 12 October 2016 establishing a discard plan for certain demersal fisheries in north-western waters.

ICES. 2016a. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

ICES. 2016b. EU request to ICES to provide Fmsy ranges for selected stocks in ICES Subareas 5 to 10 . In Report of the ICES Advisory Committee, 2016. ICES Advice 2016, Book 5, Section 5.2.3.1.

\section*{Table 1. Whiting in Divisions 7.bc,e-k. Nominal Landings (t) as reported to ICES, and total landings as used by the Working Group.}

\({ }^{\text {a Provisional data. }}{ }^{\mathrm{b}}\) French Official landings not available, not updated.

Table 2. Whiting in Divisions 7.bc-ek. Landings (t) by fleet.
\begin{tabular}{lccccccc}
\hline Fleet & BEL & FRA & \multicolumn{1}{l}{ IRL } & UK & Others & Total & \(\%\) \\
\hline OTB & 15 & 3404 & 3242 & 448 & 7 & 7115 & \(78 \%\) \\
\hline SSC & 8 & 0 & 1283 & 2 & 25 & 1317 & \(15 \%\) \\
\hline TBB & 81 & 0 & 68 & 51 & 0 & 199 & \(2 \%\) \\
\hline Other & 0 & 282 & 58 & 87 & 6 & 433 & \(5 \%\) \\
\hline & 103 & 3686 & 4650 & 588 & 37 & 9065 & \(100 \%\) \\
\hline
\end{tabular}

Table 3. Whiting in Divisions 7.b,c,e-k. Length distributions for Landings (LAN) and Discards (DIS) for 2017 by country and main fleet (Numbers in '000s) available in InterCatch.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline & \multirow[t]{2}{*}{\begin{tabular}{l}
BEL \\
TBB
\end{tabular}} & \multicolumn{4}{|l|}{FRA} & \multicolumn{4}{|l|}{IRL} & \multicolumn{4}{|l|}{UK} \\
\hline & & ОТВ & OTB & OTT & OTT & ОТВ & Отв & TBB & TBB & ОТВ & ОTB & TBB & TBB \\
\hline Lngt_cm & LAN & LAN & DIS & LAN & DIS & LAN & DIS & LAN & DIS & LAN & DIS & LAN & DIS \\
\hline 10 & & & 6 & & & & 14 & & & & 3 & & 7 \\
\hline 11 & & & 1 & & & & 42 & & & & 3 & & \\
\hline 12 & & & 3 & & 1 & & 91 & & & & 6 & & 3 \\
\hline 13 & & & 11 & & 2 & & 160 & & & & 4 & & \\
\hline 14 & & & 12 & & 2 & & 169 & & & & 10 & & 3 \\
\hline 15 & & & 20 & & 2 & & 179 & & & & 11 & & \\
\hline 16 & & & 20 & & 3 & & 204 & & & & 14 & & 3 \\
\hline 17 & & & 25 & & 2 & & 149 & & & & 8 & & \\
\hline 18 & & & 48 & & 4 & & 219 & & & & 3 & & 7 \\
\hline 19 & & & 57 & & 3 & & 96 & & & & 5 & & 3 \\
\hline 20 & & & 48 & & 3 & & 101 & & & & 13 & & 11 \\
\hline 21 & & & 26 & & 2 & & 104 & & 2 & & 4 & & 24 \\
\hline 22 & & & 33 & & 3 & & 157 & & & & 8 & & 40 \\
\hline 23 & & & 29 & & 2 & & 123 & & 2 & & 5 & & 58 \\
\hline 24 & & 0 & 48 & 0 & 3 & & 89 & & 5 & & 11 & & 68 \\
\hline 25 & & 0 & 51 & & 3 & & 90 & & 3 & 4 & 13 & & 59 \\
\hline 26 & & 18 & 40 & & 3 & & 125 & & 6 & 3 & 32 & & 48 \\
\hline 27 & 1 & 31 & 101 & 0 & 6 & & 158 & & 15 & 8 & 69 & 0 & 39 \\
\hline
\end{tabular}
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline & BEL & FRA & & & & IRL & & & & UK & & & \\
\hline & TBB & Отв & ОтВ & OTT & OTT & ОТВ & ОTB & TBB & TBB & ОТВ & Отв & TBB & TBB \\
\hline 28 & 3 & 84 & 66 & 1 & 6 & 3 & 218 & & 13 & 24 & 100 & 0 & 68 \\
\hline 29 & 4 & 184 & 103 & 2 & 7 & 3 & 286 & & 14 & 43 & 101 & 1 & 64 \\
\hline 30 & 10 & 284 & 73 & 3 & 8 & 11 & 375 & & 10 & 58 & 109 & 1 & 46 \\
\hline 31 & 16 & 394 & 50 & 5 & 7 & 108 & 304 & & 8 & 81 & 116 & 2 & 32 \\
\hline 32 & 19 & 545 & 44 & 7 & 5 & 124 & 221 & & 16 & 118 & 83 & 4 & 52 \\
\hline 33 & 21 & 538 & 41 & 8 & 3 & 189 & 90 & 0 & 12 & 107 & 49 & 6 & 48 \\
\hline 34 & 24 & 458 & 53 & 7 & 3 & 201 & 49 & 1 & 11 & 108 & 30 & 6 & 41 \\
\hline 35 & 24 & 508 & 4 & 8 & 1 & 275 & 16 & 2 & 7 & 86 & 12 & 9 & 35 \\
\hline 36 & 16 & 456 & 16 & 8 & 1 & 319 & 31 & 4 & 7 & 82 & 8 & 7 & 21 \\
\hline 37 & 13 & 320 & 1 & 7 & 0 & 366 & 14 & 5 & 4 & 54 & 5 & 8 & 11 \\
\hline 38 & 10 & 376 & 1 & 7 & & 313 & 12 & 7 & 2 & 52 & 5 & 5 & 12 \\
\hline 39 & 9 & 301 & 6 & 7 & & 304 & 10 & 7 & 1 & 60 & 2 & 7 & 2 \\
\hline 40 & 7 & 295 & 10 & 6 & 0 & 279 & 6 & 8 & 0 & 42 & 2 & 5 & 18 \\
\hline 41 & 5 & 260 & & 7 & 0 & 361 & & 10 & 1 & 34 & 1 & 5 & 10 \\
\hline 42 & 3 & 237 & & 6 & & 251 & & 9 & & 30 & & 5 & 3 \\
\hline 43 & 2 & 207 & & 5 & & 234 & & 6 & & 25 & & 2 & 2 \\
\hline 44 & 1 & 164 & & 4 & & 173 & & 6 & & 20 & 1 & 2 & \\
\hline 45 & 1 & 212 & & 5 & & 233 & & 6 & & 22 & & 4 & 1 \\
\hline 46 & 1 & 161 & & 5 & & 268 & & 6 & 1 & 14 & & 3 & \\
\hline 47 & 1 & 78 & & 4 & & 174 & & 4 & 0 & 9 & & 3 & \\
\hline 48 & 1 & 112 & & 4 & & 201 & & 3 & & 6 & & 2 & \\
\hline 49 & & 82 & & 2 & & 130 & & 2 & & 6 & & 2 & \\
\hline 50 & & 112 & & 3 & & 118 & & 2 & & 7 & & 1 & \\
\hline 51 & & 52 & & 2 & & 86 & & 1 & & 4 & & 1 & 1 \\
\hline 52 & & 88 & & 2 & & 91 & & 2 & & 4 & & 1 & \\
\hline 53 & & 26 & & 1 & & 68 & & 2 & & 4 & & 1 & \\
\hline 54 & & 25 & & 1 & & 57 & & & & 2 & & 1 & \\
\hline 55 & & 22 & & 1 & & 48 & & & & 2 & & 1 & \\
\hline
\end{tabular}
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline & \multirow[t]{2}{*}{\begin{tabular}{l}
BEL \\
TBB
\end{tabular}} & \multirow[t]{2}{*}{\begin{tabular}{l}
FRA \\
OTB
\end{tabular}} & \multirow[b]{2}{*}{ОТВ} & \multirow[b]{2}{*}{OTT} & \multirow[b]{2}{*}{OTT} & \multirow[t]{2}{*}{\begin{tabular}{l}
IRL \\
ОTB
\end{tabular}} & \multirow[b]{2}{*}{ОTB} & \multirow[b]{2}{*}{TBB} & \multirow[b]{2}{*}{TBB} & \multirow[t]{2}{*}{\begin{tabular}{l}
UK \\
ОTB
\end{tabular}} & \multirow[b]{2}{*}{OTB} & \multirow[b]{2}{*}{TBB} & \multirow[b]{2}{*}{TBB} \\
\hline & & & & & & & & & & & & & \\
\hline 56 & & 7 & & 1 & & 23 & & & & 2 & & & \\
\hline 57 & & 10 & & 0 & & 25 & & & & 1 & & & \\
\hline 58 & & 8 & & 0 & & 10 & & 2 & & 0 & & & \\
\hline 59 & & 2 & & 0 & & 11 & & & & 1 & & & \\
\hline 60 & & 2 & & 0 & & 7 & & & & & & & \\
\hline 61 & & 2 & & 0 & & 1 & & & & & & & \\
\hline 62 & & & & 0 & & 1 & & & & & & & \\
\hline 63 & & 1 & & 0 & & 1 & & & & & & & \\
\hline 64 & & & & 0 & & & & & & & & & \\
\hline 65 & & 1 & & & & 1 & & & & & & & \\
\hline 68 & & & & & & 0 & & & & & & & \\
\hline 71 & & & & 0 & & & & & & & & & \\
\hline \multicolumn{14}{|l|}{73} \\
\hline \multicolumn{14}{|l|}{75} \\
\hline \multicolumn{14}{|l|}{80} \\
\hline 90 & & & & & & & & & & & & & \\
\hline
\end{tabular}

Table 4. Whiting in Divisions 7.bc,e-k. The strong 1999 year class is distinct in both the catch and landings data, with evidence of the strong 2009 and 2013 year classes appearing at older ages. Catch numbers-at-age ('000).
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline 1999 & \multicolumn{7}{|l|}{2018} \\
\hline 0 & 7 & & & & & & \\
\hline \multicolumn{8}{|l|}{1} \\
\hline 5370.0 & 20744.1 & 25957.7 & 14662.4 & 8744.8 & 8987.8 & 6670.2 & 1498.7 \\
\hline 8176.3 & 26561.7 & 26303.7 & 12529.9 & 6122.5 & 2605.9 & 2100.9 & 2424.3 \\
\hline 8795.0 & 26105.8 & 51390.6 & 13715.2 & 5317.1 & 2049.0 & 763.1 & 627.3 \\
\hline 4568.6 & 13387.4 & 34319.6 & 24356.6 & 5968.2 & 1057.6 & 291.6 & 111.0 \\
\hline 3687.0 & 12213.5 & 11836.5 & 10634.3 & 12778.4 & 1640.7 & 227.8 & 58.1 \\
\hline 2473.8 & 27330.2 & 15052.2 & 6542.4 & 7241.9 & 6212.0 & 573.2 & 81.2 \\
\hline 1421.1 & 10663.5 & 32482.0 & 12581.9 & 5079.9 & 4819.8 & 3717.7 & 155.1 \\
\hline 5114.1 & 29760.2 & 44102.5 & 10995.4 & 4217.2 & 1750.4 & 1181.6 & 579.4 \\
\hline 1017.0 & 14791.8 & 36137.0 & 12258.9 & 5296.7 & 1407.4 & 345.4 & 325.7 \\
\hline 1650.1 & 8270.8 & 13274.5 & 6373.7 & 3290.8 & 858.5 & 214.8 & 68.4 \\
\hline 538.1 & 8045.5 & 20840.4 & 7931.2 & 2653.7 & 770.3 & 192.4 & 201.5 \\
\hline 348.0 & 4004.6 & 12591.3 & 10429.8 & 4761.1 & 1201.0 & 260.9 & 101.4 \\
\hline 737.0 & 4691.4 & 8226.7 & 8280.5 & 5464.3 & 1738.5 & 355.4 & 84.5 \\
\hline 156.0 & 5399.4 & 6661.7 & 10006.3 & 5577.9 & 1725.5 & 505.5 & 116.1 \\
\hline 739.0 & 1076.3 & 6880.1 & 7160.1 & 10810.1 & 4379.2 & 938.2 & 216.5 \\
\hline 158.7 & 13119.4 & 5727.8 & 7237.2 & 6301.1 & 7941.1 & 2032.8 & 352.8 \\
\hline 262.3 & 4167.2 & 25419.9 & 8601.1 & 7555.1 & 2619.8 & 4343.9 & 805.3 \\
\hline 1223.7 & 9891.3 & 11827.4 & 29870.3 & 5397.2 & 3145.3 & 1160.7 & 1933.0 \\
\hline 1055.7 & 6577.3 & 13369.2 & 7914.1 & 11829.3 & 2550.3 & 1437.8 & 315.5 \\
\hline 223.8 & 2215.4 & 6562.5 & 6880.8 & 3238.0 & 3646.3 & 824.0 & 496.7 \\
\hline
\end{tabular}

Table 5. Whiting in Divisions 7.bc,e-k. Catch weights-at-age (kg).
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline 1999 & \multicolumn{7}{|l|}{2018} \\
\hline 0 & 7 & & & & & & \\
\hline \multicolumn{8}{|l|}{1} \\
\hline 0.0271 & 0.1331 & 0.2216 & 0.3412 & 0.4274 & 0.4402 & 0.4963 & 0.6230 \\
\hline 0.0314 & 0.0690 & 0.2204 & 0.3955 & 0.5053 & 0.5630 & 0.5804 & 0.5868 \\
\hline 0.0315 & 0.1116 & 0.1853 & 0.3778 & 0.5293 & 0.6335 & 0.7600 & 0.7775 \\
\hline 0.0272 & 0.0965 & 0.1966 & 0.3506 & 0.5315 & 0.7069 & 0.8249 & 1.0133 \\
\hline 0.0290 & 0.0945 & 0.2114 & 0.3604 & 0.4521 & 0.6291 & 0.8306 & 1.0873 \\
\hline 0.0401 & 0.1554 & 0.2266 & 0.3612 & 0.4321 & 0.4910 & 0.5366 & 0.7846 \\
\hline 0.0198 & 0.1047 & 0.1950 & 0.3608 & 0.5010 & 0.5038 & 0.4869 & 0.6744 \\
\hline 0.0333 & 0.1235 & 0.2103 & 0.3855 & 0.5377 & 0.5878 & 0.5443 & 0.6750 \\
\hline 0.0419 & 0.1214 & 0.2014 & 0.3644 & 0.4975 & 0.6423 & 0.6088 & 0.6382 \\
\hline 0.0283 & 0.1093 & 0.2141 & 0.3859 & 0.5241 & 0.6264 & 0.7795 & 0.8298 \\
\hline 0.0257 & 0.1168 & 0.2064 & 0.3950 & 0.5492 & 0.6530 & 0.6894 & 0.9506 \\
\hline 0.0344 & 0.1190 & 0.2278 & 0.4205 & 0.5601 & 0.6793 & 0.8151 & 0.8356 \\
\hline 0.0243 & 0.1261 & 0.2393 & 0.4435 & 0.6130 & 0.8109 & 0.9538 & 1.2106 \\
\hline 0.0387 & 0.0956 & 0.2248 & 0.4607 & 0.6493 & 0.8084 & 0.9671 & 1.0881 \\
\hline 0.0533 & 0.1303 & 0.2086 & 0.3576 & 0.6002 & 0.7042 & 0.9147 & 0.8644 \\
\hline 0.0380 & 0.1420 & 0.2543 & 0.3968 & 0.5536 & 0.6621 & 0.7588 & 1.0072 \\
\hline 0.0180 & 0.1017 & 0.2199 & 0.3754 & 0.5726 & 0.7777 & 0.6711 & 0.9295 \\
\hline 0.0517 & 0.1489 & 0.2170 & 0.3584 & 0.5772 & 0.6847 & 0.7457 & 0.7836 \\
\hline 0.0531 & 0.1540 & 0.2508 & 0.3895 & 0.4996 & 0.5695 & 0.7391 & 0.8509 \\
\hline 0.0456 & 0.1463 & 0.2424 & 0.3917 & 0.5885 & 0.7063 & 0.8278 & 0.9755 \\
\hline
\end{tabular}

Table 6. Whiting in Divisions 7.bc,e-k. Q1 Stock weights-at-age (kg) from Rivard corrected annual mean catch weights.


Table 7. Whiting in Divisions 7.e-k. Summary of landings and discard data in 2017 provided to the Working Group.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|}
\hline \multicolumn{11}{|l|}{weight in tonnes} \\
\hline \multirow[t]{2}{*}{DISCARDS} & \multicolumn{2}{|l|}{Country} & 1 & 2 & 3 & 4 & 5 & 6 & 7+ & Grand Total \\
\hline & Belgium & 0.7 & 58.4 & 91.1 & 92.8 & 33.9 & 13.6 & 0.1 & 0.0 & 290.5 \\
\hline & France & 7.7 & 73.0 & 99.6 & 73.1 & 18.4 & 7.0 & 0.3 & 0.0 & 279.2 \\
\hline & Ireland & 0.0 & 31.4 & 218.5 & 229.2 & 75.2 & 4.0 & 0.0 & 0.0 & 558.3 \\
\hline & UK (England) & 1.8 & 68.5 & 152.8 & 104.1 & 24.7 & 13.3 & 0.1 & 0.0 & 365.2 \\
\hline & Other & 0.0 & 0.2 & 0.7 & 0.5 & 0.1 & 0.0 & 0.0 & 0.0 & 1.6 \\
\hline & Total & 10.2 & 231.4 & 562.7 & 499.7 & 152.4 & 37.9 & 0.5 & 0.1 & 1494.8 \\
\hline \multirow[t]{6}{*}{Landings} & Belgium & 0.0 & 0.7 & 9.0 & 27.9 & 24.5 & 34.6 & 5.7 & 2.4 & 104.8 \\
\hline & France & 0.0 & 63.9 & 507.1 & 926.6 & 547.1 & 737.3 & 262.9 & 349.0 & 3393.9 \\
\hline & Ireland & 0.0 & 14.0 & 392.8 & 1022.6 & 1070.4 & 1631.9 & 395.7 & 122.4 & 4649.9 \\
\hline & UK (England) & 0.0 & 13.6 & 114.5 & 209.1 & 104.3 & 123.4 & 14.5 & 8.1 & 587.5 \\
\hline & Other & 0.0 & 0.4 & 4.5 & 9.6 & 7.0 & 10.2 & 2.9 & 2.5 & 37.2 \\
\hline & Total & 0.0 & 92.7 & 1028.0 & 2195.9 & 1753.2 & 2537.4 & 681.7 & 484.5 & 8773.3 \\
\hline \multicolumn{11}{|l|}{Number in 000's} \\
\hline \multirow[t]{3}{*}{Discards} & Country & & 1 & 2 & 3 & 4 & 5 & 6 & 7 & Grand Total \\
\hline & Belgium & 17.2 & 412.3 & 454.7 & 327.5 & 87.8 & 30.7 & 0.1 & 0.0 & 1330.4 \\
\hline & France & 147.9 & 447.6 & 442.4 & 257.0 & 51.2 & 16.6 & 0.4 & 0.1 & 1363.1 \\
\hline
\end{tabular}
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|}
\hline \multicolumn{11}{|l|}{weight in tonnes} \\
\hline \multirow[t]{2}{*}{DISCARDS} & \multicolumn{2}{|l|}{COUNTRY} & 1 & 2 & 3 & 4 & 5 & 6 & 7+ & Grand Total \\
\hline & Ireland & 0.0 & 511.8 & 1985.1 & 1273.5 & 273.3 & 11.0 & 0.0 & 0.0 & 4054.7 \\
\hline & UK (England) & 58.3 & 504.5 & 732.4 & 375.0 & 72.8 & 30.3 & 0.1 & 0.0 & 1773.4 \\
\hline & Other & 0.4 & 1.7 & 3.6 & 2.0 & 0.3 & 0.1 & 0.0 & 0.0 & 8.1 \\
\hline & Total & 223.8 & 1877.9 & 3618.2 & 2235.0 & 485.4 & 88.7 & 0.6 & 0.1 & 8529.8 \\
\hline \multirow[t]{6}{*}{Landings} & Belgium & 0.0 & 2.4 & 23.7 & 52.9 & 36.3 & 45.8 & 6.6 & 2.2 & 170.0 \\
\hline & France & 0.0 & 233.9 & 1464.4 & 2025.3 & 933.4 & 1196.1 & 323.7 & 351.6 & 6528.2 \\
\hline & Ireland & 0.0 & 48.7 & 1047.5 & 2006.0 & 1575.2 & 2119.6 & 473.0 & 133.3 & 7403.3 \\
\hline & UK (England) & 0.0 & 51.0 & 396.6 & 542.3 & 196.9 & 181.4 & 16.5 & 7.2 & 1392.0 \\
\hline & Other & 0.0 & 1.6 & 12.0 & 19.3 & 10.8 & 14.7 & 3.5 & 2.3 & 64.3 \\
\hline & Total & 0.0 & 337.5 & 2944.3 & 4645.8 & 2752.6 & 3557.6 & 823.4 & 496.6 & 15557.8 \\
\hline
\end{tabular}

Table 8. Whiting in Divisions 7.bc,e-k. Combined survey abundance indices of age groups \(\mathbf{- 5}\) (NB: values for 2017 comprised of extended Irish survey to compensate from missing French survey).

\section*{IGFSEVHOE No/Hr}


Table 9. Whiting in Divisions 7.bc,e-k. XSA Diagnostics.

Run 1

FLR XSA Diagnostics 2019-05-15 09:41:01

CPUE data from indices

Catch data for 20 years 1999 to 2018. Ages 0 to 7 .
fleet first age last age first year last year alpha beta

1 IGFSExtendedNo/Hr \(0 \quad 5 \quad 20032018\) <NA> <NA>

Time-series weights :

Tapered time weighting not applied
Catchability analysis :

Catchability independent of size for all ages
Catchability independent of age for ages \(>5\)

Terminal population estimation
Survivor estimates shrunk towards the mean F
of the final 5 years or the 3 oldest ages.
S.E. of the mean to which the estimates are shrunk \(=1\)

Minimum standard error for population
estimates derived from each fleet \(=0.5\)
prior weighting not applied

Regression weights
year
age 2009201020112012201320142015201620172018
\(\begin{array}{lllllllllll}\text { all } & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}\)

Fishing mortalities
year
age 2009201020112012201320142015201620172018
00.0000 .0000 .0000 .0000 .0000 .0000 .0000 .0000 .0000 .000
10.0330 .0110 .0300 .0350 .0090 .0440 .0410 .1110 .1010 .042
20.2820 .1180 .0500 .0970 .1000 .1070 .2070 .2910 .4110 .255
30.3330 .3420 .1590 .1180 .2160 .2190 .3610 .6520 .5150 .627
```

4 0 . 4 7 8 0 . 4 7 5 0 . 4 1 6 ~ 0 . 2 0 4 ~ 0 . 2 4 3 ~ 0 . 4 1 3 0 . 5 2 3 0 . 5 6 9 0 . 8 6 8 ~ 0 . 5 7 7 ~
5 0 . 4 4 6 0 . 5 4 4 0 . 4 1 0 0 . 2 8 3 0 . 3 1 3 0 . 3 6 6 ~ 0 . 3 8 8 ~ 0 . 5 6 9 ~ 0 . 7 9 4 ~ 1 . 0 3 5 )
6.3330.330 0.380 0.245 0.304 0.291 0.442 0.371 0.728 0.865
70.3330.330 0.380 0.245 0.304 0.291 0.442 0.3710.7280.865

```

XSA population number (Thousand)
age
\(\begin{array}{lllllllll}\text { year } & 0 & 1 & 2 & 3 & 4 & 5 & 6 & 7\end{array}\)

200918775463814101174283592286642615821846 2010823287554308156164462451561134961122429 2011827878243059231957724271992663151360318 2012628396244415998021151483748185552810637 20131577179185521999154728862048198834322982 20145399744656317780547189231053164497421665 201549058815941718850336480229859948147102673 2016356770144836647488004115427885945237401 201727832510532954855252562528456833363717 20187310148217040293189779155690717211004

Estimated population abundance at 1st Jan 2019
age
\(\begin{array}{lllllllll}\text { year } & 0 & 1 & 2 & 3 & 4 & 5 & 6 & 7\end{array}\)

201902158193333016293615233441644496

Fleet: IGFSExtendedNo/Hr

Log catchability residuals.
year
age 2003200420052006200720082009201020112012201320142015201620172018 \(0-0.7640 .431-0.7480 .2560 .686-0.099-0.131-1.2440 .130-0.0960 .530-0.7860 .1950 .5381 .1020 .000\)

\title{
\(10.4220 .098-0.740-0.050-0.1100 .0610 .072-0.2650 .255-0.080-0.3110 .043-0.0580 .2850 .927-0.550\) \(20.5350 .471-0.8260 .361-0.199 \quad 0.120-0.3210 .0930 .1910 .212-0.102-0.3710 .1590 .6150 .273-1.210\) \(3-0.0480 .394-0.3690 .216-0.139 \quad 0.147-0.344 \quad 0.1970 .5730 .158-0.222-0.232-0.3980 .2610 .484-0.678\) \(4-0.136-0.1950 .1550 .904-0.3630 .475-0.129-0.2170 .653-0.4340 .0560 .064-0.2340 .468-0.693-0.373\) \(5-0.682-0.2540 .5530 .447-0.072 \quad 0.3930 .0270 .0330 .135-0.653-0.028-0.068 \quad 0.5380 .315-0.096-0.588\)
}

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time
\(\begin{array}{llllll}0 & 1 & 2 & 3 & 4 & 5\end{array}\)

Mean_Logq -6.6702 -6.5897-6.6200 -6.9415-7.2004-7.1118
S.E_Logq 0.44670 .44670 .44670 .44670 .44670 .4467

Terminal year survivor and F summaries:
,Age 0 Year class =2018
source
scaledWts survivors yrcls

IGFSExtendedNo/Hr 122158192018
,Age 1 Year class =2017
source
scaledWts survivors yrcls

IGFSExtendedNo/Hr 0.793192222017
fshk \(\quad 0.207 \quad 220442017\)
,Age 2 Year class =2016
source

,Age 6 Year class \(=2012\)
source
scaledWts survivors yrcls
fshk \(1 \quad 5972012\)

Table 10. Whiting in Divisions 7.b, c, e-k. Fishing mortality (F)-at-age. Fbar range is 2-5.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|}
\hline & 1 & 2 & 3 & 4 & 5 & 6 & 7+ & Fbar \({ }^{\text {- }}\) 5 \\
\hline 2009 & 0.033 & 0.282 & 0.333 & 0.478 & 0.446 & 0.333 & 0.333 & 0.385 \\
\hline 2010 & 0.011 & 0.118 & 0.342 & 0.475 & 0.544 & 0.330 & 0.330 & 0.370 \\
\hline 2011 & 0.030 & 0.050 & 0.159 & 0.416 & 0.410 & 0.380 & 0.380 & 0.259 \\
\hline 2012 & 0.035 & 0.097 & 0.118 & 0.204 & 0.283 & 0.245 & 0.245 & 0.176 \\
\hline 2013 & 0.009 & 0.100 & 0.216 & 0.243 & 0.313 & 0.304 & 0.304 & 0.218 \\
\hline 2014 & 0.044 & 0.107 & 0.219 & 0.413 & 0.366 & 0.291 & 0.291 & 0.276 \\
\hline 2015 & 0.041 & 0.207 & 0.361 & 0.523 & 0.388 & 0.442 & 0.442 & 0.370 \\
\hline 2016 & 0.111 & 0.291 & 0.652 & 0.569 & 0.569 & 0.371 & 0.371 & 0.520 \\
\hline 2017 & 0.101 & 0.411 & 0.515 & 0.868 & 0.794 & 0.728 & 0.728 & 0.647 \\
\hline 2018 & 0.042 & 0.255 & 0.627 & 0.577 & 1.035 & 0.865 & 0.865 & 0.624 \\
\hline
\end{tabular}

Table 11. Whiting in Divisions 7.b, c, e-k. Stock number-at-age (‘000).
\begin{tabular}{lllllllll}
\hline year & \(\mathbf{0}\) & \(\mathbf{1}\) & \(\mathbf{2}\) & \(\mathbf{3}\) & \(\mathbf{4}\) & \(\mathbf{5}\) & \(\mathbf{6}\) & \(\mathbf{7}\) \\
\hline 2009 & 1877546 & 381410 & 117428 & 35922 & 8664 & 2615 & 821 & 846 \\
\hline 2010 & 823287 & 554308 & 156164 & 46245 & 15611 & 3496 & 1122 & 429 \\
\hline 2011 & 827878 & 243059 & 231957 & 72427 & 19926 & 6315 & 1360 & 318 \\
\hline 2012 & 628396 & 244415 & 99802 & 115148 & 37481 & 8555 & 2810 & 637 \\
\hline 2013 & 1577179 & 185521 & 99915 & 47288 & 62048 & 19883 & 4322 & 982 \\
\hline 2014 & 539974 & 465631 & 77805 & 47189 & 23105 & 31644 & 9742 & 1665 \\
\hline 2015 & 490588 & 159417 & 188503 & 36480 & 22985 & 9948 & 14710 & 2673 \\
\hline 2016 & 356770 & 144836 & 64748 & 80041 & 15427 & 8859 & 4523 & 7401 \\
\hline 2017 & 278325 & 105329 & 54855 & 25256 & 25284 & 5683 & 3363 & 717 \\
\hline & 731014 & 82170 & 40293 & 18977 & 9155 & 6907 & 1721 & 1004 \\
\hline
\end{tabular}

Table 12. Whiting in Divisions 7.b, c, e-k. Summary table.
\begin{tabular}{|c|c|c|c|c|c|c|}
\hline & RECRUITS & TOTALBIO & TOTSPBIO & CATCH & YIELD/SSB & \(\mathrm{F}_{\text {bar }} \mathbf{2 - 5}\) \\
\hline 1999 & 2305731 & 120232 & 50358 & 25600 & 0.508 & 0.719 \\
\hline 2000 & 1365080 & 94515 & 42448 & 20044 & 0.472 & 0.702 \\
\hline 2001 & 626150 & 85886 & 50757 & 23073 & 0.455 & 0.778 \\
\hline 2002 & 719054 & 78311 & 57627 & 20976 & 0.364 & 0.577 \\
\hline 2003 & 972995 & 68259 & 45333 & 14657 & 0.323 & 0.423 \\
\hline 2004 & 990224 & 83083 & 39250 & 16669 & 0.425 & 0.405 \\
\hline 2005 & 899817 & 66943 & 40891 & 18907 & 0.462 & 0.713 \\
\hline 2006 & 774876 & 61182 & 34549 & 21691 & 0.628 & 0.770 \\
\hline 2007 & 983016 & 69901 & 29891 & 17542 & 0.587 & 0.890 \\
\hline 2008 & 1291906 & 63432 & 25827 & 8739 & 0.338 & 0.502 \\
\hline 2009 & 1877546 & 79408 & 34946 & 10673 & 0.305 & 0.385 \\
\hline 2010 & 823287 & 95221 & 49749 & 11522 & 0.232 & 0.370 \\
\hline 2011 & 827878 & 104148 & 78031 & 11452 & 0.147 & 0.259 \\
\hline 2012 & 628396 & 109381 & 84341 & 12261 & 0.145 & 0.176 \\
\hline 2013 & 1577179 & 142762 & 78155 & 14914 & 0.191 & 0.218 \\
\hline 2014 & 539974 & 119800 & 66763 & 16824 & 0.252 & 0.276 \\
\hline 2015 & 490588 & 87363 & 74356 & 19275 & 0.259 & 0.370 \\
\hline 2016 & 356770 & 71833 & 54055 & 22457 & 0.415 & 0.520 \\
\hline 2017 & 278325 & 53102 & 34901 & 16198 & 0.464 & 0.647 \\
\hline 2018 & 731014 & 48878 & 24379 & 10268 & 0.421 & 0.624 \\
\hline Geomean & 844529 & & & & & \\
\hline Mean & 952990 & 85182 & 49830 & 16687 & 0.370 & 0.516 \\
\hline Units & (Thousands) & (Tonnes) & (Tonnes) & (Tonnes) & & \\
\hline
\end{tabular}

Table 13. Whiting in Divisions 7.b, c, e-k. Management options table.
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline Fmult & Catch18 & Land18 & Dis18 & FCatch18 & FLand18 & FDis18 & SSB19 \\
\hline & & & & & NA & NA & 58900 \\
\hline 0 & 0 & 0 & 0 & 0 & NA & NA & 39236 \\
\hline 0.1 & 1261 & 770 & 491 & 0.06236 & 0.04394 & 0.01841 & 38151 \\
\hline 0.2 & 2471 & 1503 & 968 & 0.12472 & 0.08789 & 0.03683 & 37114 \\
\hline 0.3 & 3634 & 2201 & 1433 & 0.18707 & 0.13183 & 0.05524 & 36123 \\
\hline 0.4 & 4753 & 2867 & 1886 & 0.24943 & 0.17577 & 0.07366 & 35175 \\
\hline 0.5 & 5828 & 3502 & 2326 & 0.31179 & 0.21972 & 0.09207 & 34268 \\
\hline 0.6 & 6864 & 4108 & 2756 & 0.37415 & 0.26366 & 0.11049 & 33399 \\
\hline 0.7 & 7861 & 4686 & 3174 & 0.4365 & 0.3076 & 0.1289 & 32567 \\
\hline 0.8 & 8821 & 5239 & 3583 & 0.49886 & 0.35155 & 0.14731 & 31770 \\
\hline 0.9 & 9748 & 5767 & 3981 & 0.56122 & 0.39549 & 0.16573 & 31005 \\
\hline 1 & 10641 & 6271 & 4370 & 0.62358 & 0.43943 & 0.18414 & 30270 \\
\hline 1.1 & 11503 & 6755 & 4749 & 0.68593 & 0.48338 & 0.20256 & 29566 \\
\hline 1.2 & 12336 & 7217 & 5119 & 0.74829 & 0.52732 & 0.22097 & 28889 \\
\hline 1.3 & 13141 & 7660 & 5481 & 0.81065 & 0.57126 & 0.23939 & 28238 \\
\hline 1.4 & 13919 & 8085 & 5835 & 0.87301 & 0.61521 & 0.2578 & 27612 \\
\hline 1.5 & 14672 & 8492 & 6180 & 0.93536 & 0.65915 & 0.27621 & 27010 \\
\hline 1.6 & 15400 & 8882 & 6518 & 0.99772 & 0.70309 & 0.29463 & 26431 \\
\hline 1.7 & 16105 & 9257 & 6848 & 1.06008 & 0.74704 & 0.31304 & 25873 \\
\hline 1.8 & 16789 & 9617 & 7172 & 1.12244 & 0.79098 & 0.33146 & 25335 \\
\hline 1.9 & 17451 & 9963 & 7488 & 1.18479 & 0.83492 & 0.34987 & 24817 \\
\hline \multicolumn{8}{|l|}{Additional Catch Options} \\
\hline Basis20 & Catch20 & Land20 & Dis & FCatch20 & FLand20 & FDis20 & SSB21 \\
\hline FMSY & 6481 & 3885 & 2597 & 0.35085 & 0.24724 & 0.10361 & 33720 \\
\hline FMSY & 4157 & 2513 & 1644 & 0.22 & 0.15 & 0.06 & 35680 \\
\hline \multicolumn{8}{|l|}{Lower} \\
\hline FMSY & 8104 & 4826 & 3277 & 0.45 & 0.32 & 0.13 & 32365 \\
\hline \multicolumn{8}{|l|}{Upper} \\
\hline \(\mathrm{F}=0\) & 0 & 0 & 0 & 0 & NA & NA & 39236 \\
\hline
\end{tabular}
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline Fmult & Catch18 & Land18 & Dis18 & FCatch18 & FLand18 & FDis18 & SSB19 \\
\hline \(\mathrm{F}=\mathrm{Fpa}\) & 13006 & 7586 & 5420 & 0.8 & 0.56376 & 0.23624 & 28347 \\
\hline \(\mathrm{F}=\mathrm{Flim}\) & 16762 & 9603 & 7159 & 1.12 & 0.78926 & 0.33074 & 25356 \\
\hline Blim & 17216 & 9841 & 7376 & 1.16249 & 0.8192 & 0.34328 & 25000 \\
\hline Bpa & 4960 & 2990 & 1970 & 0.26127 & 0.18412 & 0.07715 & 35000 \\
\hline Btrigger & 4960 & 2990 & 1970 & 0.26127 & 0.18412 & 0.07715 & 35000 \\
\hline \(F=F 2019\) & 10641 & 6271 & 4370 & 0.62358 & 0.43943 & 0.18414 & 30270 \\
\hline Min FMSY & 6017 & 3613 & 2404 & 0.323 & 0.22762 & 0.09538 & 34109 \\
\hline \begin{tabular}{l}
Max \\
FMSY
\end{tabular} & 11286 & 6633 & 4653 & 0.67 & 0.47215 & 0.19785 & 29743 \\
\hline Stable SSB & 19001 & 10760 & 8240 & 1.33858 & 0.9433 & 0.39529 & 23615 \\
\hline -15\% TAC & 30999 & 16306 & 14692 & 3.06844 & 2.16232 & 0.90611 & 14941 \\
\hline \begin{tabular}{l}
Stable \\
TAC
\end{tabular} & 38530 & 19184 & 19346 & 5.00105 & 3.52423 & 1.47682 & 10173 \\
\hline \[
\begin{aligned}
& +15 \% \\
& \text { TAC }
\end{aligned}
\] & 47392 & 22062 & 25331 & 9.07506 & 6.39518 & 2.67988 & 5290 \\
\hline
\end{tabular}

Input units are thousands and kg output in tonnes.

Table 14. Whiting in divisions 7.b, c, e-k. Input values for the catch forecast.

\section*{Whiting in the Celtic Sea (7.b,c, e-k), WGCSE 2018, COMBSEX}
\(F_{\text {bar }}\) age range: 2-5
nyears +1
\begin{tabular}{llllllllllll}
\hline Age & N & M & Mat & PF & PM & SWt & Sel & CWt & DSel & DCWt \\
\hline 1 & 215817 & 0.86 & 0 & 0 & 0 & 0.076 & 0.007 & 0.272 & 0.077 & 0.132 \\
\hline 2 & 33330 & 0.65 & 1 & 0 & 0 & 0.178 & 0.086 & 0.348 & 0.233 & 0.179 \\
\hline 3 & 16293 & 0.5 & 1 & 0 & 0 & 0.295 & 0.433 & 0.483 & 0.165 & 0.233 \\
\hline 4 & 6151 & 0.43 & 1 & 0 & 0 & 0.456 & 0.617 & 0.61 & 0.055 & 0.317 \\
\hline 5 & 3344 & 0.4 & 1 & 0 & 0 & 0.598 & 0.769 & 0.713 & 0.03 & 0.36 \\
\hline 6 & 1644 & 0.38 & 1 & 0 & 0 & 0.72 & 0.629 & 0.797 & 0.026 & 0.576 \\
\hline 7 & 791 & 0.36 & 1 & 0 & 0 & 0.87 & 0.638 & 0.902 & 0.017 & 0.555 \\
\hline
\end{tabular}
nyears +2
\begin{tabular}{lllllllllll}
\hline Age & N & M & Mat & PF & PM & SWt & Sel & CWt & DSel & DCWt \\
\hline 1 & 249330 & 0.86 & 0 & 0 & 0 & 0.076 & 0.007 & 0.272 & 0.077 & 0.132 \\
\hline 2 & 83591 & 0.65 & 1 & 0 & 0 & 0.178 & 0.086 & 0.348 & 0.233 & 0.179 \\
\hline 3 & 12463 & 0.5 & 1 & 0 & 0 & 0.295 & 0.433 & 0.483 & 0.165 & 0.233 \\
\hline 4 & 5292 & 0.43 & 1 & 0 & 0 & 0.456 & 0.617 & 0.61 & 0.055 & 0.317 \\
\hline 5 & 1985 & 0.4 & 1 & 0 & 0 & 0.598 & 0.769 & 0.713 & 0.03 & 0.36 \\
\hline 6 & 972 & 0.38 & 1 & 0 & 0 & 0.72 & 0.629 & 0.797 & 0.026 & 0.576 \\
\hline 7 & 846 & 0.36 & 1 & 0 & 0 & 0.87 & 0.638 & 0.902 & 0.017 & 0.555 \\
\hline
\end{tabular}
nyears +3
\begin{tabular}{llllllllllll}
\hline Age & N & M & Mat & PF & PM & SWt & Sel & CWt & DSel & DCWt \\
\hline 1 & 249330 & 0.86 & 0 & 0 & 0 & 0.076 & 0.007 & 0.272 & 0.077 & 0.132 \\
\hline 2 & 96572 & 0.65 & 1 & 0 & 0 & 0.178 & 0.086 & 0.348 & 0.233 & 0.179 \\
\hline 3 & 31258 & 0.5 & 1 & 0 & 0 & 0.295 & 0.433 & 0.483 & 0.165 & 0.233 \\
\hline 4 & 4048 & 0.43 & 1 & 0 & 0 & 0.456 & 0.617 & 0.61 & 0.055 & 0.317 \\
\hline 5 & 1708 & 0.4 & 1 & 0 & 0 & 0.598 & 0.769 & 0.713 & 0.03 & 0.36 \\
\hline 6 & 577 & 0.38 & 1 & 0 & 0 & 0.72 & 0.629 & 0.797 & 0.026 & 0.576 \\
\hline 7 & 633 & 0.36 & 1 & 0 & 0 & 0.87 & 0.638 & 0.902 & 0.017 & 0.555 \\
\hline
\end{tabular}

Input units are thousands and kg output in tonnes.

Table 15. Whiting in Divisions 7.e-k. The detailed output for the status quo F forecast by age group.

NYears+1
\begin{tabular}{llcccccccccc}
\hline Age & F & CatchNos & Yield & DF & DCatchNos & DYield & StockNos & Biomass & SSNos & SSB \\
\hline & & & & & & & 971263 & 21594 & & \\
\hline 1 & 0.008 & 1609 & 437 & 0.081 & 10728 & 1421 & 215817 & 16496 & 0 & 0 \\
\hline 2 & 0.09 & 2499 & 870 & 0.244 & 4579 & 818 & 33330 & 5943 & 33330 & 5943 \\
\hline 3 & 0.452 & 3610 & 1744 & 0.172 & 2500 & 582 & 16293 & 4805 & 16293 & 4805 \\
\hline 4 & 0.644 & 2080 & 1268 & 0.057 & 502 & 159 & 6151 & 2804 & 6151 & 2804 \\
\hline 5 & 0.803 & 1356 & 967 & 0.032 & 247 & 89 & 3344 & 1999 & 3344 & 1999 \\
\hline 6 & 0.657 & 639 & 509 & 0.027 & 53 & 31 & 1644 & 1184 & 1644 & 1184 \\
\hline 7 & 0.666 & 295 & 266 & 0.018 & 40 & 22 & 791 & 688 & 791 & 688 \\
\hline Total & 0.497 & 12088 & 6061 & 0.126 & 18649 & 3122 & 1121899 & 57566 & 61553 & 17423 \\
\hline
\end{tabular}

NYears+2
\begin{tabular}{llcccccccccc}
\hline Age & F & CatchNos & Yield & DF & DCatchNos & DYield & StockNos & Biomass & SSNos & SSB \\
\hline & & & & & & & 971263 & 21594 & & \\
\hline 1 & 0.008 & 1859 & 505 & 0.081 & 12394 & 1641 & 249330 & 19057 & 0 & 0 \\
\hline 2 & 0.09 & 6269 & 2182 & 0.244 & 11484 & 2051 & 83591 & 14904 & 83591 & 14904 \\
\hline 3 & 0.452 & 2761 & 1334 & 0.172 & 1912 & 445 & 12463 & 3676 & 12463 & 3676 \\
\hline 4 & 0.644 & 1790 & 1091 & 0.057 & 432 & 137 & 5292 & 2412 & 5292 & 2412 \\
\hline 5 & 0.803 & 805 & 574 & 0.032 & 147 & 53 & 1985 & 1187 & 1985 & 1187 \\
\hline 6 & 0.657 & 378 & 301 & 0.027 & 32 & 18 & 972 & 700 & 972 & 700 \\
\hline 7 & 0.666 & 316 & 285 & 0.018 & 43 & 24 & 846 & 736 & 846 & 736 \\
\hline Total & 0.497 & 14178 & 6272 & 0.126 & 26444 & 4369 & 1199008 & 66319 & 105149 & 23615 \\
\hline
\end{tabular}

NYears+3
\begin{tabular}{llllllllllll}
\hline Age & F & CatchNos & Yield & DF & DCatchNos & DYield & StockNos & Biomass & SSNos & SSB \\
\hline & & & & & & & 971263 & 21594 & & \\
\hline 1 & 0.008 & 1859 & 505 & 0.081 & 12394 & 1641 & 249330 & 19057 & 0 & 0 \\
\hline 2 & 0.09 & 7242 & 2520 & 0.244 & 13267 & 2369 & 96572 & 17219 & 96572 & 17219 \\
\hline 3 & 0.452 & 6925 & 3346 & 0.172 & 4796 & 1117 & 31258 & 9219 & 31258 & 9219 \\
\hline 4 & 0.644 & 1369 & 835 & 0.057 & 331 & 105 & 4048 & 1845 & 4048 & 1845 \\
\hline 5 & 0.803 & 692 & 494 & 0.032 & 126 & 45 & 1708 & 1021 & 1708 & 1021 \\
\hline 6 & 0.657 & 224 & 179 & 0.027 & 19 & 11 & 577 & 416 & 577 & 416 \\
\hline 7 & 0.666 & 236 & 213 & 0.018 & 32 & 18 & 633 & 551 & 633 & 551 \\
\hline Total & 0.497 & 18547 & 8092 & 0.126 & 30965 & 5306 & 1228655 & 72975 & 134796 & 30271 \\
\hline
\end{tabular}





Figure 1. Irish landings for the main gear types in 1995-2016, along with annual average.

TBB



Figure 2. Whiting in 7.b, c, e-k (Celtic Sea). 2018 length compositions (raised numbers 000's) of French, UK and Irish Landings (LAN) and Discards (DIS) available in InterCatch for the main fleets.


Figure 3. Whiting in 7.b, c, e-k (Celtic Sea), annual Landings (green) and Discards (red) by age composition.

Whiting.27.7b-ce-k
Rivard Corrected stock weights


Figure 4. Whiting in 7.b, c, e-k (Celtic Sea). Rivard corrected stock weights-at-age.


Figure 5. 2018 Annual proportions-at-age of Discard nos in the Stock (above); Discard nos in the Catch (middle) and Catch nos in the Stock (below) from the assessment.

\section*{IGFSExtendedNo/Hr}


Figure 6. Whiting in 7.b, c, e-k (Celtic Sea). Pairwise scatterplots for the \(\log\) numbers-at-age for the IGFSEVHOE combined survey index.


Figure 7. Whiting in 7.e-k (Celtic Sea). Mean log standardized plots of indices by year class (top panel) and by year (lower panel).


Figure 8. Whiting in 7.b, c, e-k (Celtic Sea). Log fleet catchability residuals bubble plots.

Whiting VIIbk XSA


Figure 9. Whiting in 7.b, c, e-k (Celtic Sea). Retrospective analysis.

Whiting VIIbk XSA
Fishing mortality


Figure 10. Whiting in 7.b, c, e-k (Celtic Sea). Fishing mortality-at-age.


Figure 11. Whiting in 7.b, c, e-k (Celtic Sea). Stock summary.

Landings yield 2020


SSB 2021


Figure 12. Whiting in Divisions 7.b, c, e-k. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.

\section*{Annex 1: List of participants}
\begin{tabular}{llll}
\hline Name & Institute & Country & E-mail \\
\hline Mikel Aristegui & Marine Institute & Ireland & Mikel.Aristegui@Marine.ie \\
\hline Ewen Bell & Cefas & UK & ewen.bell@cefas.co.uk \\
\hline Lynda Blackadder & Marine Science Scotland & Scotland, UK & lynda.blackadder@scotland.gsi.gov.uk \\
\hline Katie Boyle & Marine Scotland-Science & Scotland, UK & Katie.Boyle@gov.scot \\
\hline Rui Catarino & International Council for the & Denmark & rui.catarino@ices.dk \\
\hline Paul Coleman & Marine Institute & Ireland & Paul.coleman@marine.ie \\
\hline Helen Dobby & Marine Scotland-Science & Scotland, UK & h.dobby@marlab.ac.uk \\
\hline
\end{tabular}
\begin{tabular}{llll} 
Jennifer Doyle & Marine Institute & Ireland & Jennifer.doyle@marine.ie
\end{tabular}
\begin{tabular}{lll} 
Mickael Drogou Ifremer & France & Mickael.Drogou@ifremer.fr
\end{tabular}
\begin{tabular}{llll}
\hline \begin{tabular}{l} 
Timothy Earl \\
Chair
\end{tabular} & Cefas & UK & timothy.earl@cefas.co.uk \\
\hline Simon Fischer & Cefas & UK & simon.fischer@cefas.co.uk \\
\hline Helen Holah & Marine Laboratory & Ucotland, UK & Helen.Holah@scotland.gsi.gov.uk \\
\hline \begin{tabular}{l} 
Kieran Hyder \\
by \\
correspondence
\end{tabular} & Cefas & Kieran.hyder@cefas.co.uk \\
\hline Andrzej Jaworski & Marine Scotland-Science & Scotland, UK & a.jaworski@marlab.ac.uk \\
\hline \begin{tabular}{l} 
Vladimir Khlivnoi \\
by \\
correspondence \\
of Marine Fisheries and \\
Oceanography(PINRO)
\end{tabular} & \begin{tabular}{l} 
Cefas
\end{tabular} & khlivn@pinro.ru \\
\hline \begin{tabular}{l} 
Vladimir \\
Laptikovsky
\end{tabular} & \begin{tabular}{l} 
Agri-food and Biosciences Institute \\
(AFBI)
\end{tabular} & Northern & Ireland, UK
\end{tabular}
\begin{tabular}{|c|c|c|c|}
\hline Name & Institute & Country & E-mail \\
\hline Sara-Jane Moore by correspondence & Marine Institute & Ireland & Sara-jane.moore@marine.ie \\
\hline \begin{tabular}{l}
Sofie Nimmegeers \\
Chair
\end{tabular} & Research Institute for agriculture, fisheries and food (ILVO) & Belgium & sofie.nimmegeers@ilvo.vlaanderen.be \\
\hline Lisa Readdy by correspondence & Cefas & UK & lisa.readdy@cefas.co.uk \\
\hline Marianne Robert & Ifremer & France & Marianne.robert@ifremer.fr \\
\hline Marta Quinzan & Cefas & UK & Marta.quinzan@cefas.co.uk \\
\hline Stephen Shaw & Cefas & UK & stephen.shaw@cefas.co.uk \\
\hline David Stokes & Marine Institute & Ireland & david.stokes@marine.ie \\
\hline Pia Suchert & Agri-food and Biosciences Institute (AFBI) & Northern Ireland, UK & Pia.Schuchert@afbini.gov.uk \\
\hline Katie Thomas & Marine Institute & Ireland & Katie.Thomas@marine.ie \\
\hline Bart Vanelslander & Research Institute for agriculture, fisheries and food (ILVO) & Belgium & bart.vanelslander@ilvo.vlaanderen.be \\
\hline Jonathan White & \begin{tabular}{l}
Demersal fisheries \\
Fisheries Ecosystems Advisory Services \\
Marine Institute
\end{tabular} & Ireland & Jonathan.white@marine.ie \\
\hline
\end{tabular}

\section*{Annex 2: Stock Annexes}

The table below provides an overview of the WGCSE Stock Annexes. Stock Annexes for other stocks are available on the ICES website Library under the Publication Type "Stock Annexes". Use the search facility to find a particular Stock Annex, refining your search in the left-hand column to include the year, ecoregion, species, and acronym of the relevant ICES expert group.
\begin{tabular}{|c|c|c|c|}
\hline Stock ID & Stock name & Last updated & Link \\
\hline anf.27.3a46 & Anglerfish (Lophius budegassa, Lophius piscatorius) in subareas 4 and 6, and in Division 3.a (North Sea, Rockall and West of Scotland, Skagerrak and Kattegat) & October 2019 & Anglerfish 3.a46 \\
\hline bss.27.4bc7d-h & Seabass (Dicentrarchus labrax) in divisions 4.b-c, 7.a, and 7.d-h (central and southern North Sea, Irish Sea, English Channel, Bristol Channel, and Celtic Sea) & March 2018 & Sea bass 47 \\
\hline cod.27.7e-k & Cod (Gadus morhua) in divisions 7.e-k (eastern English Channel and southern Celtic Seas) & March 2016 & Cod 7.e-k \\
\hline cod.27.7a & Cod (Gadus morhua) in Division 7.a (Irish Sea) & March 2017 & Cod 7.a \\
\hline cod.27.6b & Cod (Gadus morhua) in Division 6.b (Rockall) & May 2013 & Cod 6.b \\
\hline cod.27.6a & Cod (Gadus morhua) in Division 6.a (West of Scotland) & March 2019 & Cod 6.a \\
\hline gug-celt & Grey gurnard in Subarea 6 and Divisions 7.a-c and e-k & March 2014 & Grey gurnard \\
\hline had.27.7b-k & Haddock (Melanogrammus aeglefinus) in divisions 7.b-k (southern Celtic Seas and English Channel) & May 2017 & Haddock 7.b-k \\
\hline had.27.7a & Haddock (Melanogrammus aeglefinus) in Division 7.a (Irish Sea) & May 2019 & Haddock 7.a \\
\hline had.27.6b & Haddock (Melanogrammus aeglefinus) in Division 6.b (Rockall) & \[
\begin{aligned}
& \text { September } \\
& 2019
\end{aligned}
\] & Haddock 6.b \\
\hline had.27.46a20 & Haddock (Melanogrammus aeglefinus) in Subarea 4, Division 6.a and Subdivision 20 (North Sea, West of Scotland, Skagerrak) & May 2009 & Haddock 6.a \\
\hline lez.27.4a6a & Megrim (Lepidorhombus ssp.) in divisions 4.a and 6.a (northern North Sea, West of Scotland) & May 2016 & Megrim 4a6a \\
\hline
\end{tabular}
\begin{tabular}{|c|c|c|c|}
\hline Stock ID & Stock name & Last updated & Link \\
\hline nep.fu. 11 & Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 11 (West of Scotland, North Minch) & May 2016 & Nephrops FU11 \\
\hline nep.fu. 12 & Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 12 (West of Scotland, South Minch) & May 2016 & Nephrops FU12 \\
\hline nep.fu. 13 & Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 13 (West of Scotland, the Firth of Clyde and Sound of Jura) & May 2017 & Nephrops FU13 \\
\hline nep.fu. 14 & Norway lobster (Nephrops norvegicus) in Division 7.a, Functional Unit 14 (Irish Sea, East) & \[
\begin{aligned}
& \text { September } \\
& 2018
\end{aligned}
\] & Nephrops FU14 \\
\hline nep.fu. 15 & Norway lobster (Nephrops norvegicus) in Division 7.a, Functional Unit 15 (Irish Sea, West) & May 2018 & Nephrops FU15 \\
\hline nep.fu. 16 & Norway lobster (Nephrops norvegicus) in divisions 7.b-c and 7.j-k, Functional Unit 16 (west and southwest of Ireland, Porcupine Bank) & March 2013 & Nephrops FU16 \\
\hline nep.fu. 17 & Norway lobster (Nephrops norvegicus) in Division 7.b, Functional Unit 17 (west of Ireland, Aran grounds) & May 2016 & Nephrops FU17 \\
\hline nep.fu. 19 & Norway lobster (Nephrops norvegicus) in divisions 7.a, 7.g, and 7.j, Functional Unit 19 (Irish Sea, Celtic Sea, eastern part of southwest of Ireland) & October 2019 & Nephrops FU19 \\
\hline nep.fu. 2021 & Norway lobster (Nephrops norvegicus) in divisions 7.g and 7.h, functional units 20 and 21 (Celtic Sea) & October 2019 & Nephrops FU2021 \\
\hline nep.fu. 22 & Norway lobster (Nephrops norvegicus) in divisions 7.g and 7.f, Functional Unit 22 (Celtic Sea, Bristol Channel) & May 2018 & Nephrops FU22 \\
\hline nep.fu. 2324 & Norway lobster (Nephrops norvegicus) in divisions 8.a and 8.b, functional units 23-24 (northern and central Bay of Biscay) & & Not available \\
\hline ple.27.7bc & Plaice (Pleuronectes platessa) in divisions 7.b-c (West of Ireland) & April 2013 & Plaice 7.bc \\
\hline ple.27.7h-k & Plaice (Pleuronectes platessa) in divisions 7h-k (Celtic Sea South, southwest of Ireland) & May 2014 & Plaice 7.h-k \\
\hline ple.27.7fg & Plaice (Pleuronectes platessa) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea) & May 2018 & Plaice 7.fg \\
\hline ple.27.7e & Plaice (Pleuronectes platessa) in Division 7.e (western English Channel) & April 2016 & Plaice 7.e \\
\hline ple.27.7a & Plaice (Pleuronectes platessa) in Division 7.a (Irish Sea) & May 2019 & Plaice 7.a \\
\hline sol.27.7bc & Sole (Solea solea) in divisions 7.b and 7.c (West of Ireland) & April 2013 & Sole 7.bc \\
\hline
\end{tabular}
\begin{tabular}{|c|c|c|c|}
\hline Stock ID & Stock name & Last updated & Link \\
\hline sol.27.7h-k & Sole (Solea solea) in divisions 7.h-k (Celtic Sea South, Southwest of Ireland) & May 2014 & Sole 7.h-k \\
\hline sol.27.7fg & Sole (Solea solea) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea) & May 2019 & Sole 7.fg \\
\hline sol.27.7e & Sole (Solea solea) in Division 7.e (western English Channel) & May 2017 & Sole 7.e \\
\hline sol.27.7a & Sole (Solea solea) in Division 7.a (Irish Sea) & May 2019 & Sole 7.a \\
\hline whg.27.7b-ce-k & Whiting (Merlangius merlangus) in divisions 7.b-c and 7.e-k (southern Celtic Seas and eastern English Channel) & June 2019 & Whiting 7.bc,e-k \\
\hline whg.27.7a & Whiting (Merlangius merlangus) in Division 7.a (Irish Sea) & May 2017 & Whiting 7.a \\
\hline whg.27.6b & Whiting (Merlangius merlangus) in Division 6.b (Rockall) & May 2013 & Whiting 6.b \\
\hline whg.27.6a & Whiting (Merlangius merlangus) in Division 6.a (West of Scotland) & May 2017 & Whiting 6.a \\
\hline
\end{tabular}

\section*{Annex 3: Working documents}

The following working documents were presented to WGCSE in 2019. They are found in full below in the following pages:

WD1: Review of changes in the reference points of bss.27.4bc7ad-h. Arni Magnusson, ICES.
WD2: Update of reference points 2019 for seabass in ICES divisions 4.b,c and 7.a, d-h. Lisa Readdy, Cefas.
WD3: Maturity-at-age estimates for Irish Demersal Stocks in 6.a, 7.a and 7.bgj between 2004-2018. Sara-Jane Moore and Hans Gerritsen, Marine Institute, Galway.
WD4: Pollack life-history parameters for MYAS project. Hans Gerritsen and Katie Thomas, Marine Institute, Galway.

WD5: Revision of the LPUE abundance index for the seabass in areas 4.b,c and 7.a, d-h. Mathieu Woillez \({ }^{1}\), Mickael Drogou \({ }^{1}\) and Alain Laurec \({ }^{2} .{ }^{1}\) Ifremer. \({ }^{2}\) ENSAR Department Halieutique.
WD6: Nephrops in the North Minch (FU11): results of the 2019 UWTV Survey and catch options for 2020. Katie Boyle, Lynda Blackadder and Helen Dobby, Marine Scotland Science.

WD7: Nephrops in the South Minch (FU12): results of the 2019 UWTV survey and catch options for 2020. Katie Boyle, Lynda Blackadder and Helen Dobby, Marine Scotland Science.

WD8: Nephrops in the Clyde (FU13): results of the 2019 UWTV survey and catch options for 2020. Katie Boyle, Lynda Blackadder and Helen Dobby, Marine Scotland Science.

WD9: Western Irish Sea Nephrops Grounds (FU15) 2019 UWTV Survey Report and catch options for 2020. Mathieu Lundy \({ }^{1}\), Peter McCorriston \({ }^{1}\), Ian McCausland \({ }^{1}\), Keith Erskine \({ }^{1}\), Katie Lilley \({ }^{1}\), Gary Heaney \({ }^{1}\),Jim McArdle \({ }^{1}\), Aaron Buick \({ }^{1}\), Jessica Graham \({ }^{1}\), Charlotte Reeve \({ }^{3}\) and Jennifer Doyle \({ }^{2}\). \({ }^{1}\) Fisheries and Aquatic Ecosystems Branch, Agri-Food \& Biosciences Institute. \({ }^{2}\) Fisheries Ecosystems Advisory Services, Marine Institute. \({ }^{3}\) Centre for Environment, Fisheries and Aquaculture Science, Lowestoft.
WD10: 012019 FU2021 SISP datacheck. Mikel Aristegui-Ezquibela and Jennifer Doyle, Marine Institute.
WD11: UWTV camera calibration test, Marine Institute. Mikel Aristegui-Ezquibela, Marine Institute.
WD12: SIAMISS Estimates of Anglerfish Biomass in subareas 4 and 6 for 2019. Elisa Barreto \({ }^{1}\), Liz Clarke \({ }^{1}\), Gerald McAllister \({ }^{1}\), James Dooley \({ }^{1}\), Ruadhan Gillespie-Mules \({ }^{1}\), Eoghan Kelly \({ }^{2}\), Hans Gerritsen \({ }^{2}\). \({ }^{1}\) Marine Scotland Science and Marine Institute, Ireland.

\title{
Review of changes in the reference points of bss.27.4bc7ad-h
}

\author{
Arni Magnusson
}

24 June 2019

\section*{Background}

Seabass in Divisions 4.b-c, 7.a, and 7.d-h (central and southern North Sea, Irish Sea, English Channel, Bristol Channel, and Celtic Sea) is a category 1.2 stock, assessed by WGCSE using the SS3 model.

The basis of this review is the WGCSE WD 03 Update of reference points 2019 for seabass in ICES Divisions 4.b,c and 7.a, \(d\)-h, the WGCSE draft report, model files on the Sharepoint data folder, and the draft advice sheet. For the purposes of the review, a TAF repository was created (2019_bss.27.4bc7ad-h_review) using the SS3 model output file Report.sso found in the Sharepoint data folder.

Changes were made to the input data in this year's assessment, Landings per Unit of Effort (LPUE) and recreational catches. This causes changes in the estimated stock size and fishing mortalities.

The main reason to revise the reference points is that \(\mathrm{B}_{\text {loss }}\) in the assessment has changed and all other reference points hinge on \(B_{\text {loss }}\) for this stock. In last year's assessment \(B_{\text {loss }}\) was \(\mathrm{SSB}_{2016}=9618 \mathrm{t}\), but in this year's assessment \(\mathrm{B}_{\text {loss }}\) is \(\mathrm{SSB}_{2018}=10313 \mathrm{t}\), an increase of \(7 \%\). The effect on reference point values is described below.

\section*{Changes in the assessment}

Two changes were made to the input data:

1 LPUE indices are now calculated based on 0.99 kg delta-GLM threshold instead of 0.05 kg .

2 Recreational catches were adjusted after the LPUE change, to fulfill the assumption that recreational \(F\) was constant in 1985-2014.

\section*{Changes in the reference points}
\begin{tabular}{lccc}
\hline Reference point & 2018 & 2019 & Change \\
\hline \(\mathrm{MSY} \mathrm{B}_{\text {trigger }}=\mathrm{B}_{\text {pa }}\) & 13465 & 14439 & \(+7 \%\) \\
\(\mathrm{~F}_{\mathrm{MSY}}\) & 0.203 & 0.1713 & \(-16 \%\) \\
\(\mathrm{~B}_{\text {lim }}=\mathrm{B}_{\text {loss }}\) & 9618 & 10313 & \(+7 \%\) \\
\(\mathrm{~F}_{\text {lim }}\) & 0.295 & 0.254 & \(-14 \%\) \\
\(\mathrm{~F}_{\text {pa }}\) & 0.211 & 0.1815 & \(-14 \%\) \\
\hline
\end{tabular}
\(\mathrm{B}_{\text {lim }}\) is set to \(\mathrm{B}_{\text {loss }}\), and \(\mathrm{B}_{\mathrm{pa}}\) is calculated as \(1.4 \times \mathrm{B}_{\text {lim }}\).
\(\mathrm{F}_{\text {MSY }}\) and \(\mathrm{F}_{\text {lim }}\) are estimated with EqSim, and \(\mathrm{F}_{\mathrm{pa}}\) is calculated as \(\mathrm{F}_{\text {lim }} / 1.4\).

\section*{Comments}
- The changes in the assessment stem from a consensus in the WG that it was incorrect to use 0.05 kg as a delta-GLM threshold for calculating the LPUE indices. The WG decided to fix this error in the analysis for the 2019 assessment.
- The changes in the reference points seem to be justified and correctly calculated.
- In the EqSim analysis, \(\mathrm{F}_{\text {MSY }}\) and \(\mathrm{F}_{\text {lim }}\) are reduced to maintain the stock above a somewhat higher \(\mathrm{B}_{\mathrm{lim}}\) than was used in last year's advice.
- The WD, report, and draft advice sheet have reference point tables showing \(\mathrm{F}_{\text {MSY }}=0.1713, \mathrm{~F}_{\text {lim }}=0.254, \mathrm{Fpa}=0.1815\). In the WD text, they are also specified with higher precision, but in the advice sheet these reference points should perhaps be rounded to \(\mathrm{F}_{\text {MSY }}=0.171, \mathrm{~F}_{\text {lim }}=0.25\), \(\mathrm{Fpa}=0.182\).
- In the WD, there is a minor typo where \(\mathrm{B}_{\mathrm{lim}}\) is specified as 10303 t in the text, but in all other places it is 10313 t .

\title{
Update of reference points 2019 for seabass in ICES Divisions 4.b,c and 7.a, d-h. \\ Lisa Readdy \\ Cefas
}

\section*{Introduction}

In 2019 a corrected set of input data was provided for the Landings per Unit of Effort series, When including the new series the vector of recreational catch also required an update in order to maintain the assumption of constant fishing mortality for the timeseries 1985 to 2014. This impacted on the output estimation of biomass, total mortality and recruitment changing the perception of the stock. Due to these updates in input data, reference points were recalculated to reflect the change in the perception of the stock. This document sets out the process in defining reference points for the Northern stock of bass, bss.27.4.b,c \& 7.a,d-h. The approach taken was to run the analysis with the 2019 update assessment using the stock synthesis framework 3.24 u (Methot \& Wetzel, 2013. The final proposed reference points are provided in table 3.

\section*{Materials and Methods}

\section*{Data}

The output files from the recent 2019 accepted assessment using stock synthesis was used to calculate reference points (Table 1). These files were produced during the assessment working group for the Celtic Seas Ecoregion.

\section*{Methods}

The same software, EqSim, as used in 2018 (ICES, 2018) was used to calculate the update reference points (Table 1) following the standard ICES guidance for category 1 and 2 stocks (ICES, 2017). An \(R\) scripts was used to transform the stock synthesis output into the format required for the EqSim software. The same R script as that used during WKBass 2018 was updated and used to produce the reference points

\section*{Settings}

Natural mortality (M) of 0.24 as accepted at the 2018 benchmark and used in the accepted assessment was used in the calculation of reference points.

The range of ages used \(\mathrm{F}_{\text {bar }}\) is the same as that used in the accepted assessment and forecast, this includes the range of ages 4 to 15 years.

With the recent implementation of management measures, increase in minimum conservation reference size, for this stock the number of years used in the reference point calculation was set at 4 years. This is an increase on that used during WKBass which was only 2 years.

Other biological input data remains as a 10 year timeseries 2009 to 2018.
Although the three stock recruitment relationships were considered, the segmented regression stockrecruitment relationship was selected using the same rational as that described in the recent benchmarks (ICES, 2016 and ICES, 2018) for bass. While the stock recruitment relationship used in the assessment is Beverton and Holt, the steepness parameter is fixed at 0.999 as accepted at the benchmark. This setting gives a relationship with no clear signal between stock size and recruitment.

Recruitment for this stock is influenced by environmental conditions such as water temperature and any stock-recruitment relationship is difficult to detect under these conditions.

\section*{Results}

Stock recruitment relationship and calculation of biomass Precautionary Approach (PA) reference points.
The weighted stock recruitment fits are presented in Figure 1. When the three stock recruitment (S-R) relationships where used, Ricker S-R was weighted the highest with it being selected \(76 \%\) of the time. Most of the difference between each of the S-R relationships are below the level of SSB where there is no available data, SSB below 10, 000 tonnes. This difference is also true for the weighted BevertonHolt and Segmented regression, with the segmented regression being selected \(75 \%\) of the time. Because most of the difference occurs where data are not available this supports the decision made to just use the segmented regression Figure 2.

Given the S-R relationship and following the ICES guidance this stock could be considered and a potential type 1 or type 5 stock.
- Type 1: Spasmodic stocks - stocks with occasional large year classes.
- Type 5: Stocks showing no evidence of impaired recruitment or with no clear relation between stock and recruitment (no apparent S-R signal).

Given the outcome of the assessment and available evidence the stock remains as a type 5 which was also agreed at WKBass therefore \(\mathrm{B}_{\text {lim }}\) is set to \(\mathrm{B}_{\text {loss }}\) and the Segmented regression was recalculated with the breakpoint at \(\mathrm{B}_{\mathrm{lim}}\) (figure 3).

With \(\mathrm{B}_{\text {lim }}\) set to \(\mathrm{B}_{\text {loss }}\) (10 303 tonnes), Bpa could then be calculated at follows.
- \(\mathrm{B}_{\mathrm{pa}}=\) Blim*1.4
- \(B_{p a}=B \lim \times \exp (1.645 \times \sigma)\) where \(\sigma\) is the uncertainty estimated from the model.

As the assessment underestimates the uncertainty, not all uncertainty is accounted for in the model, and the uncertainty is less than the default the first option was used giving a final value of 14439 tonnes.

\section*{Calculation of fishing motality PA reference points, \(F_{p a}\) and \(F_{\text {lim. }}\).}
\(F_{\text {lim }}\) and \(\mathrm{F}_{\mathrm{pa}}\) was estimated using the EqSim software was used with the settings for Fcv, Fphi and MSY Btrigger set to zero (i.e. no assessment/advice error and no MSY \(B_{\text {triger }}\) used). The S-R used was that selected from the previous exercise, a segmented regression with the breakpoint equal to \(\mathrm{B}_{\mathrm{lim}}\).

Flim is estimated as the fishing mortality that, at equilibrium from a long-term stochastic projection, leads to a \(50 \%\) probability of having SSB above Blim. Flim was estimated to be \(\mathbf{0 . 2 5 4 1 2 0 3}\), and \(\mathrm{F}_{\mathrm{pa}}\) is estimated to be \(\mathbf{0 . 1 8 1 5 1 4 5}\) based on the following equation \(\left[\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\mathrm{lim}} \times 1.4^{-1}\right]\). The Alternative calculation, \(\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} \times \exp \times-1.645 \times \sigma\), was not used as not all uncertainty is taken into account and the value is less than the default setting for \(\sigma\).

\section*{Calculation of MSY reference points, \(F_{M S Y}\), Flower, Fupper and MSY Btrigger.}

Initially, \(F_{\text {Ms }}\) is calculated as the fishing mortality that maximises median long-term yield in stochastic simulations under constant F exploitation (i.e. without MSY \(\mathrm{B}_{\text {trigger }}\) ). Using the same simulation method with the inclusion of assessment/advice error default values: \(\mathrm{Fcv}=0.212\), \(\mathrm{Fphi}=0.423\) from WKMSYREF3 (ICES, 2015) and shown in table 1. \(\mathrm{F}_{\mathrm{MSY}}=0.196997\) and is thus above \(\mathrm{F}_{\mathrm{pa}}=0.1815145\), see figure 4 and 5 . In such a case, \(F_{\text {Msy }}\) is reduced to \(F_{\text {pa }}\) (i.e. \(F_{\text {Msy }}\) cannot exceed \(F_{p a}\) ).

ICES defines MSY \(B_{\text {trigger }}\) as the 5 th percentile of the distribution of SSB when fishing at \(\mathrm{F}_{\text {MSY }}\). However if the stock has not been fished at \(F_{M S Y}\), as in this case, table 2 , then MSY \(B_{\text {trigger }}\) is set to \(B_{\text {pa }}\).

For this final run, assessment/advice error were included using the same default values and MSY \(\mathrm{B}_{\text {trigger }}\) was Set to 14439 tonnes. As shown in figure 6, EqSim output \(\mathrm{F}_{\text {p. } 05}\) (fishing mortality that gives \(5 \%\) probability of SSB below \(B_{\text {lim }}\) ) equals 0.1712511 . As \(F_{\text {MSY }}\) estimated in the first run is above \(F_{p .05}\), then \(F_{\text {MSY }}\) is further reduced to \(F_{p .05}, 0.1712511\). F Fower was adjusted accordingly to take account of \(F_{\text {MSY }}\) rescaling to \(F_{p .05}\). \(F_{\text {upper }}\) was set to the same value as \(F_{\text {p. } 05}\) as \(F_{\text {MSY }}\) was estimated as greater than \(F_{p .05}\).

Tables
Table 1. Settings and inputs for the calculation of reference points.
\begin{tabular}{|c|c|c|}
\hline Data and parameters & Setting & Comments \\
\hline SSB-recruitment data & Full data series (years classes 1985-2018) & \\
\hline Exclusion of extreme values (option extreme.trim) & No & \\
\hline Trimming of R values & No & \\
\hline Mean weights and proportion mature; natural mortality & These parameters are constant in SS•, the same values used. & \\
\hline Exploitation pattern & 2015-2018 & \\
\hline Assessment error in the advisory year. CV of F & 0.212 & Default value calculated from 5 stocks in WKMSYREF3 (ICES, 2015) \\
\hline Autocorrelation in assessment error in the advisory year & 0.423 & Default value calculated from 5 stocks in WKMSYREF3 (ICES, 2015) \\
\hline
\end{tabular}

Table 2. \(\mathrm{F}_{\text {bar }}\) for last 10 years above \(\mathrm{F}_{\mathrm{MSY}}\)
\begin{tabular}{|r|rrrrrrrrrr|}
\hline Year & 2009 & 2010 & 2011 & 2012 & 2013 & 2014 & 2015 & 2016 & 2017 & 2018 \\
\hline Fbar \(^{2}\) & 0.201 & 0.233 & 0.217 & 0.236 & 0.271 & 0.218 & 0.197 & 0.121 & 0.099 & 0.078 \\
\hline
\end{tabular}

Table 3: Summary table of proposed updated reference points using the WGCSE 2019 accepted assessment.
\begin{tabular}{|c|c|c|}
\hline Stock & \multicolumn{2}{|l|}{Seabass in ICES Divisions 4.b,c and 7.a, d-h.} \\
\hline PA Reference points & Value 2019 & Rational \\
\hline Blim & 10313 & Lowest observed SSB (Type 5 S-R relationship) \\
\hline \(\mathrm{Bpa}_{\text {р }}\) & 14439 & \(\mathrm{Blim}^{\times} 1.4\) \\
\hline Flim & 0.254 & In equilibrium gives a \(50 \%\) probability of \(\mathrm{SSB}>\mathrm{B}_{\mathrm{lim}}\) \\
\hline \(\mathrm{Fpa}_{\text {p }}\) & 0.1815 & \(\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} / 1.4\) \\
\hline MSY Reference point & & \\
\hline Fmsy & 0.1713 (0.197) & Reduce as \(\mathrm{F}_{\mathrm{ms}}>\mathrm{FFA}_{\mathrm{Pa}}>\mathrm{Fp}_{\text {p }}\). 05 \\
\hline FMSY lower & 0.142 &  \\
\hline FmsY upper & 0.1713 & \begin{tabular}{l}
Reduce as Fmsy \(>\) Fpa \(>\) Fp. 05 \\

\end{tabular} \\
\hline MSY Btrigger & 14439 & \\
\hline
\end{tabular}

Figures


Figure 1. Stock recruitment model fit. The x-axis corresponds to SSB (tonnes). Left: Weighted S-R for Ricker, Beverton-Holt and Segmented regression. Right: Weighted S-R for Beverton-Holt and Segmented regression.

Predictive distribution of recruitme for Bass47


Figure 2. Stock recruitment model fit. Segmented regression, the x-axis corresponds to SSB (tonnes).


Figure 3. Stock recruitment model fit. Segmented regression fixing breakpoint at \(\mathrm{B}_{\text {lim }}=\mathrm{B}_{\text {loss }}\), the x -axis corresponds to SSB (tonnes).


Figure 4. EqSim summary plot without MSY \(\mathrm{B}_{\text {trigger. }}\). Panels a to c : historic values (dots) median (solid black) and \(90 \%\) intervals (dotted black) recruitment, SSB and landings for exploitation at fixed values of F. Panel c also shows mean landings (red solid line). Panel d shows the probability of SSB<Blim (red), \(S S B<B_{p a}\) (green) and the cumulative distribution of FMSY based on yield as landings (brown).


Figure 5. Left) Eqsim median landings yield curve with estimated reference points without MSY \(B_{\text {trigger }}\). Blue lines: \(\mathrm{F}_{\text {MSY }}\) estimate (solid) and range at \(95 \%\) of maximum yield (dotted). Green lines: F (5\%) estimate (solid) and range at 95\% of yield implied by F (5\%) (Dot-ted). Right) EqSim median SSB curve with estimated reference points without MSY \(B_{\text {trigger. }}\). Blue dots: lower and upper SSB corresponding to lower and upper \(\mathrm{F}_{\text {msr. }}\).


Figure 6. Eqsim median landings yield curve with estimated reference points with Btrigger. Blue lines: \(\mathrm{F}_{\text {msy }}\) estimate (solid) and range at \(95 \%\) of maximum yield (dotted). Green lines: F (5\%) estimate (solid) and range at \(95 \%\) of yield implied by F (5\%) (Dotted).

\section*{References}

ICES. 2015. Report of the joint ICES-MYFISH workshop to consider the basis for FMSY ranges for all stocks (WKMSYREF3), 17-21 November 2014, Charlottenlund, Denmark.. ICES CM 2014/ACOM:64. 156 pp.
ICES. 2016. Report of the Second Inter-Benchmark Protocol for Sea Bass in the Irish Sea, Celtic Sea, English Channel, and Southern North Sea (IBP-Bass2). By correspondence. ICES CM 2016/ ACOM:31. 191 pp.
ICES. 2017. ICES fisheries management reference points for category 1 and 2 stocks. ICES Advice 2017, Book 12, Section 12.4.3.1.
ICES. 2018. Report of the Benchmark Workshop on Seabass (WKBASS), 20-24 February 2017 and 21-23 February 2018, Copenhagen, Denmark. ICES CM 2018/ACOM:44. 259 pp.
Methot, R. D., and Wetzel, C. R. 2013. Stock synthesis: A biological and statistical framework for fish stock assessment and fishery management. Fisheries Research, 142: 86-99.

\title{
Maturity-at-age estimates for Irish Demersal Stocks in 6.a, 7.a and 7.bgj between 2004-2018
}

\author{
Sara-Jane Moore and Hans Gerritsen \\ Marine Institute \\ Galway \\ Ireland
}

\section*{Introduction}

This document provides maturity-at-age estimates for stocks assessed by the WGCSE and WGBIE. All data are obtained on surveys and commercial sampling carried out by the Marine Institute.

\section*{Methods}

Data was used from the Marine Institute Q1 Biological sampling programme (2010-2018), AtSea Observer programme (2010-2018), Irish Anglerfish and megrim survey (2016-2018), the Irish beam trawl Ecosystem survey (2016-2018) and the MI Biological sampling survey (20042009). Proportions mature-at-age were estimated by constructing a matrix containing the sample numbers by age, sex and maturity state (mature/immature) at each length class. Unsexed individuals (usually small fish with undeveloped gonads) were assigned in equal numbers to both sexes. This Age-Sex-Maturity-Length Key (ASMLK) was applied to the lengthfrequency data to estimate the proportions mature-at-age for either sex and both sexes combined. Any gaps in the ASMLK were filled in using a multinomial model (Gerritsen et al., 2006).

\section*{Results}

Figure 1 shows that for most stocks there are no clear trends in the L50 over time. Estimates for cod in area \(7(\operatorname{cod} 7)\) varied from around 40 cm to 60 cm , however the sample sizes for this stock were generally very low at the start of the time-series; in recent years the estimates are were quite variable (around 40 cm ). Sole in 7 also exhibited variable estimates in recent years. Plaice in area 7 (ple 7) had an outlying estimate for 2013 but this was estimated with low precision. Because overall there was no clear evidence of trends in maturity over time for any stock, data from all years (2004-2018) were combined. Table 1. Shows the estimated
proportions mature-at-age. For the cod stocks, the proportion of mature 2-year-olds is somewhat higher than that the proportions used by the working group. For other ages the estimates are very similar. For haddock in 7.b-k the Irish estimates are slightly lower for 2-year-olds and in agreement for the other ages. For haddock in 7.a the Irish estimates are similar to those used by WGCSE, 2018 for all ages. For haddock in 6.a the Irish estimates for age 2 were higher than the proportions used by the WGNSSK working group. For megrim, the Irish estimates were very close for females of ages 2 to 4 , for ages 5 to 8 the Irish estimates were somewhat lower than those used by the WGBIE working group. Estimated proportions mature for plaice and sole were also slightly lower than those used by the working group. For whiting in 7.b-k, the Irish maturity estimates are broadly in agreement with the ogives used by the working group, for the other whiting stocks the Irish estimates are considerably higher for the 0-group and similar for older fish.

\section*{Discussion}

Some (relatively minor) differences were found between the ogives used by the working groups and the current findings. Because Irish sampling generally does not cover the full extent of the stocks, it is difficult to determine whether the Irish estimates are unbiased. It is possible that the lack of full spatial coverage can explain some of the differences.

\section*{References}

Gerritsen, H.D., Armstrong, M.J., Allen, M., McCurdy, W.J. and Peel, J.A.D., 2003. Variability in maturity and growth in a heavily exploited stock: whiting (Merlangius merlangus L.) in the Irish Sea. J. Sea Res., 49(1): 69-82.
Gerritsen, H.D., McGrath, D. and Lordan, C., 2006. A simple method for comparing age-length keys reveals significant regional differences within a single stock of haddock (Melanogrammus aeglefinus). ICES J. Mar. Sci., 63(3): 1096-1100.
ICES, 2018. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 9-18 May, 2018, ICES CM 2018/ACOM:13


Figure 1. Length at \(50 \%\) maturity ( \(\mathrm{L} 50 ; \mathrm{cm}\) ) for females by stock and year.

Table 1. Estimated proportions mature (sample numbers in brackets) by stock, sex and age. Maturity ogives used by the WG are also given.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline Stock & Sex/WG & 1 & 2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10+ \\
\hline \multirow[t]{2}{*}{cod 7} & F & 0.02 (760) & 0.60 (942) & 0.95 (120) & 1.00 (20) & 1.00 (3) & 1.00 (3) & 1.00 (2) & & & \\
\hline & M & 0.01 (922) & 0.75 (1494) & 0.98 (133) & 1.00 (14) & 1.00 (2) & & & & & \\
\hline cod 7.a & WGCSE & 0 & 0.64 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 \\
\hline \(\operatorname{cod} 7 . e-k\) & WGCSE & 0 & 0.39 & 0.87 & 0.93 & 1 & 1 & 1 & 1 & 1 & 1 \\
\hline \multirow[t]{3}{*}{had 7.b-k} & F & 0.01 (384) & 0.89 (888) & 0.98 (714) & 0.98 (247) & 1.00 (107) & 1.00 (58) & 1.00 (47) & 1.00 (21) & 1.00 (10) & 1.00 (3) \\
\hline & M & 0.27 (493) & 0.79 (726) & 0.89 (482) & 0.89 (172) & 1.00 (81) & 1.00 (30) & 0.96 (19) & 1.00 (15) & 1.00 (4) & 1.00 (1) \\
\hline & WGCSE & 0 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 \\
\hline \multirow[t]{3}{*}{had 7.a} & F & 0.02 (154) & 0.78 (198) & 0.96 (129) & 1.00 (5) & 1.00 (5) & & & & & \\
\hline & M & 0.14 (112) & 0.72 (183) & 0.87 (125) & 1.00 (3) & 1.00 (1) & & & & & \\
\hline & WGCSE & 0 & 0.72 & 0.99 & 1 & 1 & 1 & 1 & 1 & 1 & 1 \\
\hline \multirow[t]{3}{*}{had 6.a} & F & 0.05 (17) & 0.91 (192) & 0.82 (204) & 0.83 (168) & 0.94 (31) & 0.96 (64) & 0.98 (49) & 1.00 (35) & 0.91 (24) & 1.00 (5) \\
\hline & M & 0.05 (35) & 0.75 (150) & 0.67 (132) & 0.72 (80) & 0.94 (12) & 0.71 (18) & 0.65 (12) & 0.34 (11) & 0.43 (7) & \\
\hline & WGNSSK & 0 & 0 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 \\
\hline \multirow[t]{3}{*}{mgw 78} & F & 0.10 (14) & 0.25 (534) & 0.67 (1096) & 0.87 (840) & 0.88 (539) & 0.88 (372) & 0.84 (192) & 0.85 (141) & 0.92 (125) & 1.00 (1) \\
\hline & M & 0.66 (15) & 0.35 (580) & 0.54 (699) & 0.69 (387) & 0.71 (234) & 0.75 (176) & 0.85 (139) & 0.90 (74) & 0.94 (31) & \\
\hline & WGHMM & 0.04 & 0.21 & 0.6 & 0.9 & 0.98 & 1 & 1 & 1 & 1 & 1 \\
\hline \multirow[t]{2}{*}{ple 7} & F & 0.00 (13) & 0.14 (222) & 0.45 (720) & 0.65 (547) & 0.78 (406) & 0.94 (164) & 0.92 (98) & 0.86 (50) & 0.93 (18) & 0.98 (28) \\
\hline & M & 0.00 (14) & 0.31 (249) & 0.57 (518) & 0.72 (380) & 0.81 (208) & 0.87 (108) & 0.87 (52) & 0.91 (39) & 0.83 (14) & 1.00 (9) \\
\hline ple 7.a & WGCSE & 0 & 0.24 & 0.57 & 0.74 & 0.93 & 1 & 1 & 1 & 1 & 1 \\
\hline ple 7.fg & WGCSE & 0 & 0.26 & 0.52 & 0.86 & 1 & 1 & 1 & 1 & 1 & 1 \\
\hline \multirow[t]{2}{*}{sol 7} & F & 0.00 (2) & 0.17 (40) & 0.47 (402) & 0.65 (698) & 0.87 (473) & 0.93 (274) & 0.96 (174) & 0.98 (100) & 0.95 (82) & 0.96 (139) \\
\hline & M & & 0.22 (22) & 0.45 (81) & 0.51 (127) & 0.59 (96) & 0.71 (132) & 0.70 (118) & 0.76 (113) & 0.69 (73) & 0.78 (164) \\
\hline sol 7.fg & WGCSE & 0 & 0.14 & 0.45 & 0.88 & 0.98 & 1 & 1 & 1 & 1 & 1 \\
\hline \multirow[t]{3}{*}{whg 7.b-k} & F & 0.29 (564) & 0.96 (661) & 0.98 (392) & 0.99 (172) & 1.00 (56) & 1.00 (10) & 1.00 (2) & 1.00 (1) & & \\
\hline & M & 0.49 (618) & 0.82 (516) & 0.95 (347) & 0.85 (159) & 0.80 (54) & 1.00 (16) & 1.00 (2) & 1.00 (1) & & \\
\hline & WGCSE & 0 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 \\
\hline \multirow[t]{3}{*}{whg 7.a} & F & 0.11 (295) & 0.92 (281) & 0.99 (144) & 1.00 (22) & 1.00 (4) & & & & & \\
\hline & M & 0.23 (239) & 0.77 (148) & 0.74 (48) & 1.00 (9) & 1.00 (5) & & & & & \\
\hline & WGCSE & 0 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 \\
\hline \multirow[t]{3}{*}{whg \(6 . a\)} & F & 0.44 (63) & 0.90 (166) & 0.93 (167) & 0.92 (109) & 0.96 (43) & 1.00 (13) & 1.00 (4) & 1.00 (2) & & 0.00 (1) \\
\hline & M & 0.54 (77) & 0.68 (136) & 0.48 (119) & 0.66 (54) & 0.79 (13) & 0.64 (12) & 0.72 (6) & 1.00 (1) & & \\
\hline & WGCSE & 0 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 \\
\hline
\end{tabular}

\title{
Pollack life-history parameters for MYAS project
}

\author{
Hans Gerritsen, Katie Thomas \\ 26 November 2018
}

\section*{MYDAS}

The MYDAS project https://github.com/laurieKell/mydas requires realistic life-history parameters for each of the case-study stocks. By default these are obtained from http://www.fishbase.org but the quality of these parameters are difficult to judge. In regard to Pollack the Marine Institute, Ireland (MI) has a reasonable amount of data available from surveys, observer trips and port sampling. Age data are available for the landings data for 2016 and 2017 and also for a number of surveys.

\section*{Length frequency of the landings}


The largest observed fish is 103 cm . That can inform us about Linf. If the growth curve flattens off in the older fish, we would expect the largest fish to be a couple of standard deviations above Linf, therefore, Linf would be assumed to be above 80 cm .
Note: There is a proportion of fish below 30 cm . (This is due to the fish originating from a single sample and could well be species mis-identification)

\section*{Length-frequency by gear}


Most of the landings are from gillnets, other gears seem to catch relatively more fish below 50 cm suggesting the gillnets may have a higher L50 for gear selectivity (perhaps they operate in areas where smaller fish are not available). There is no obvious difference in the selectivity of the larger fish (with gillnets we expect a domeshaped selection). Perhaps otter trawls and seines catch a few more fish 770 cm relative to gillnets.

Note: There is also the issue of availability; younger fish simply do not seem to be available to the commercial fishery (probably due to remaining inshore during the juvinile stages). Beam trawls seem to catch the smallest pollack.

\section*{Biological data}

Age data is available fish from the landings in 2016-17 from the IAMS (anglerfish) surveys 2016-17, also from IBES (beam trawl) 2016-17 and from IGFS (IBTS) 2016-17. Some individual weights and maturity data are available from other IGFS surveys.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|}
\hline \multicolumn{9}{|c|}{\# Groups: dataType [?]} \\
\hline \# & & dataType & dataSourc & total & aged & sex & mat & wt \\
\hline \#\# & & <fct> & <fct> & <int> & <int> & <int> & <int> & <int> \\
\hline \#\# & 1 & 1 Landings & Lan2015 & 4 & 0 & 4 & 4 & 4 \\
\hline \#\# & 2 & 2 Landings & Lan2016 & 438 & 436 & 0 & NA & 438 \\
\hline \# & 3 & 3 Landings & Lan2017 & 646 & 645 & 0 & NA & 646 \\
\hline \#\# & & 4 Landings & Lan2018 & 623 & 0 & 0 & NA & 621 \\
\hline \#\# & 5 & 5 Survey & IAMS2016 & 36 & 36 & 36 & 36 & 36 \\
\hline \#\# & & 6 Survey & IAMS2017 & 39 & 39 & 39 & 39 & 39 \\
\hline
\end{tabular}
\begin{tabular}{lllrrrrr} 
\#\# & 7 & Survey & IBES2016 & 1 & 1 & 1 & 1 \\
\#\# & 8 & Survey & IBES2017 & 3 & 2 & 3 & 3 \\
\#\# & 9 & Survey & IGFS2010 & 38 & 0 & 38 & 38 \\
\#\# & 10 & Survey & IGFS2011 & 38 & 0 & 38 & 38 \\
\#\# & 11 & Survey & IGFS2012 & 23 & 0 & 23 & 23 \\
\#\# & 12 & Survey & IGFS2013 & 82 & 0 & 82 & 82 \\
\#\# & 13 & Survey & IGFS2014 & 42 & 0 & 41 & 42 \\
\#\# & 14 & Survey & IGFS2015 & 9 & 0 & 9 & 9 \\
\#\# & 15 & Survey & IGFS2016 & 24 & 24 & 24 & 24 \\
\#\# & 16 & Survey & IGFS2017 & 18 & 18 & 18 & 18 \\
\hline
\end{tabular}

\section*{Growth}

Fitting a VBGF to the raw age data gives the parameters below. Note: that there may be some bias due to length-stratified sampling.
\begin{tabular}{lrrr} 
\#\# & Linf & K & t0 \\
\#\# & 84.6191408 & 0.1941284 & -0.9428006
\end{tabular}

These parameters are very similar to those given by fishbase ( \(\operatorname{Linf}=85.6, \mathrm{k}=0.19, \mathrm{t} 0=0\) ) except for t 0 . The fishbase parameters do not fit very well, due to the difference in \(t 0\).


Conclusion: do not use the fishbase parameters but those based on MI data: \(\operatorname{Linf}=84.6 ; \mathrm{k}=0.19 ; \mathrm{t}=-0.94\)

\section*{Length-weight}
```


## 

## Call:

## lm(formula = log(indWt) ~ log(lenCls/10))

## 

## Coefficients:

## (Intercept) log(lenCls/10)

## -4.541 2.984

```

Compare MI data to fishbase


The France fishbase data appears incorrect and is perhaps due to weights being gutted weight as opposed to total weight. The MI data seems reliable.
Conclusion: the suggested final length-weight parameters are: \(\mathrm{a}=0.0107 ; \mathrm{b}=2.98\)

\section*{Maturity}

There seems to be an increasing number of mature fish up to 60 cm when all data is included


It appears that all the immature fish are from the Q4 surveys. At that time of year it is difficult to distinguish between virgin and recovered fish. The data from the beggining of the year (spawning time) suggests nearly \(100 \%\) mature fish in the catches


Conclusion: all fish caught in the spring surveys are mature; immature fish may not be available to the surveys. The assumed age of first maturity is 3 .

\section*{Total mortality and selectivity}

If we apply the age length key data to the length frequency data we can get the numbers-at-age in the landings. We can then use this to see assess mortality from one age to the next. This is a rough measure of the total mortality \((\mathrm{Z})\). If you substract the natural mortality from this, you get a ballpark figure of F .


The slope in the plot above is fitted over the average landings numbers at ages 5 to 9 over the two years. The slope is -0.72 which suggests that F may be around 0.52 , which is high; quite possibly above Fmsy

A bonus in estimating Z is that we extrapolate Z over the younger ages to see how many you would expect if selectivity and Z were the same for all ages:


Conclusion: it looks like the age at \(50 \%\) selectivity is somewhere between 2 and 3 .

\section*{Thompson-Bell yield-per recruit}

We can estimate yield per recruit using the approach by Thompson and Bell (1934). For now assuming \(M=0.2\), knife-edge maturity at age 3 and selectivity as reported above.


F01 can be used as a proxy for Fmsy. The F=Z-0.2 estimate of 0.52 is well above that. However F0.1 is quite conservative and F is below Fmax. The YPR is very flat-topped.

\section*{Summary}

Growth parameters: \(\operatorname{Linf}=84.6 ; \mathrm{k}=0.19 ; \mathrm{t}=-0.94\)
Length-weight parameters: \(\mathrm{a}=0.0107 ; \mathrm{b}=2.98\)
Maturity: knife edge at age 3 (?)
Selectivity: A50 between ages 2 and 3
Z: 0.72
F01: 0.29

\title{
Working Document on the revision of the LPUE abundance index for the seabass in areas 4b,c and 7a,d-h
}

\author{
Mathieu Woillez, Mickael Drogou and Alain Laurec
}

13/06/2019

\section*{Correction of an error in the computation}

A bug was found in the computation of the French LPUE abundance index for seabass in areas \(4 b, c\) and \(7 a, d-h\). For reminder, the LPUE abundance index is the product of 1) an index modelling the proportion of days at sea where positive catch were recorded and 2) an index modelling the average catch for those positive days. Both indices relied on a threshold that defined what a positive value is. The threshold accepted at the benchmark was 0.99 kg as mentioned in the report and the many annexes on the LPUE index (ICES, 2018).

However, in WKBASS2017/2018 and in WGCSE 2018, 0.05 kg was used as a threshold to define the first index, while 0.99 kg was used to define the second one. This was a mistake. This inconsistency between both indices that must be ultimately combined was thus corrected for WGCSE 2019. 0.99 kg is now used for both indices.

Figure 1 compares both series used for the assessments in WGCSE 2018 and 2019. The main differences occurs before 2009. The explanation comes from the fact that false positive values occurred in the log-book dataset over this period.


Figure 1: LPUE abundance indices used for WGCSE 2018 and 2019.

\section*{Heterogeneity in the very low catch values of seabass over time}

Figure 2 corresponds to the overall histogram (all years from 2001 up to 2018) of the seabass catch values. Rounding is observed with spike appearing at \(1 \mathrm{~kg}, 2 \mathrm{~kg}, 3 \mathrm{~kg}\)... Decimal values also exist and are frequent for values below 1 kg . Those very low catch values reported in the log books may not be indicating of real catches of seabass, but rather of false positive values, i.e. zeroes. Indeed, 0.01 kg or 0.02 kg catches can be considered as a nonsense, as the minimum landing size of 36 cm corresponds to a seabass of 0.5 kg . A very low catch value may come from the division of low seabass catch for a given trip by the number of days of the trip.


Figure 2: left) Overall histogram of the seabass catch values in areas \(4 b, c\) and \(7 a, d-h\). Right) Zoom over the range \([0 ; 9 \mathrm{~kg}]\) of the overall histogram of the seabass catch values from the left.

There is indeed some heterogeneity in the logbook dataset with a shift occurring between 2008 and 2009 (Table 1). It is feared that many of the very low catches (many are declared at 0.1 kg ) are false positives. Before 2009, proportions of those false positive values are between \(3 \%\) and \(9 \%\). After 2009, proportions of false positive values are pretty close to \(0 \%\). In order to homogenize the whole, a threshold defining what a positive catch is, was required. A threshold of 0.99 kg was considered correctly and reported in every previous working document, accepted by the review group, but mistake happens.

Table 1: Yearly proportions of positive ( \(>0.99 \mathrm{~kg}\) ), epsilon ( \(>0.05 \mathrm{~kg}\) and <= 0.99 kg ) and zero ( < \(=0.05\) kg ) seabass catch values in areas \(\mathbf{4 b}, \mathrm{c}\) and \(7 \mathrm{a}, \mathrm{d}-\mathrm{h}\).
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline Variables & 2000 & 2001 & 2002 & 2003 & 2004 & 2005 & 2006 & 2007 & 2008 & 2009 & 2010 & 2011 & 2012 & 2013 & 2014 & 2015 & 2016 & 2017 & 2018 \\
\hline false positives & 0.09 & 0.07 & 0.08 & 0.08 & 0.08 & 0.05 & 0.03 & 0.03 & 0.04 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 \\
\hline true positives & 0.43 & 0.46 & 0.43 & 0.4 & 0.44 & 0.38 & 0.37 & 0.39 & 0.42 & 0.28 & 0.27 & 0.26 & 0.29 & 0.25 & 0.26 & 0.25 & 0.24 & 0.24 & 0.24 \\
\hline zeroes & 0.48 & 0.47 & 0.49 & 0.47 & 0.48 & 0.58 & 0.60 & 0.59 & 0.54 & 0.72 & 0.73 & 0.74 & 0.71 & 0.75 & 0.74 & 0.74 & 0.76 & 0.76 & 0.76 \\
\hline
\end{tabular}

\section*{Sensitivity analysis on the threshold defining what a positive values is}

A sensitivity analysis on the threshold value was conducted. No threshold for the definition of positive days or catches is perfect, so that various values from 0.05 to 9.99 were tested. Figure 3 shows that the results are quite robust after 2009 in areas \(4 b, c\) and \(7 \mathrm{a}, \mathrm{d}-\mathrm{h}\). The main differences appear before 2009, where the log-book dataset is polluted by many very low values (Table 1).

The threshold value was thus kept equal to the threshold accepted during the benchmark, i.e. at 0.99 kg , and used for the WGCSE 2019 assessment.


Figure 3: Sensitivity analysis on the threshold defining what a positive value is when computing the LPUE abundance index for WGCSE 2019.

\section*{Assessment model fit to the new LPUE time series and consequences}

Fit of the assessment models to the LPUE time series are illustrated in the Figure 4. The fit is better with the new LPUE abundance index. The new expected LPUE index is always within the observed uncertainty range and closer to the average observed estimates in the assessment done during WGCSE 2019. This was not the case in the assessment done during WGCSE 2018. In addition, the linear relationship between the expected LPUE index and the observed LPUE index is improved in the assessment done during WGCSE 2019.


Figure 4: Top) Assessment model fit to the LPUE abundance index used in the WGCSE 2019. Bottom) Assessment model fit to the LPUE abundance index used in the WGCSE 2018.

So the revised series should be considered valid for the current assessment done during WGCSE 2019.

\section*{Reference}

ICES. 2018. Report of the Benchmark Workshop on Seabass (WKBASS), 20-24 February 2017 and 21-23 February 2018, Copenhagen, Denmark. ICES CM 2018/ACOM:44. 259 pp.

\title{
Nephrops in the North Minch (FU 11): results of the 2019 UWTV Survey and catch options for 2020
}

\author{
Katie Boyle, Lynda Blackadder \& Helen Dobby
}

September 2019
Marine Scotland Science
Aberdeen

\section*{Introduction}

Since the early 1990s Marine Scotland Science (MSS) has carried out annual underwater television surveys (UWTV) targeting Nephrops norvegicus which have been used to estimate Nephrops abundance based on burrow counts. UWTV surveys for the estimation of Nephrops abundance reduce the problems associated with traditional trawl surveys that arise from variability in burrow emergence of Nephrops. UWTV surveys are targeted at known areas of mud, sandy mud and muddy sand in which Nephrops construct burrows.

UWTV surveys provide an absolute estimate of stock abundance in numbers which are used to estimate harvest ratios, defined as the ratio of total catch to total abundance. The landed proportion of the catch converted to biomass has been used by ICES to provide TAC advice.

UWTV surveys in the North Minch FU 11 have been conducted since 1994 (no surveys in 1995 and 1997). The survey takes place annually in May/June and also covers other Functional Units to the West of Scotland (FU 12 and FU 13) and in the North Sea (FU 7 and 34) and in 2019 FU10. The main objectives for the 2019 UWTV survey are listed below:
1) Obtain estimates of the abundance and distribution of Nephrops burrow complexes in each FU
2) Use the video footage to record the occurrence of other benthic fauna as well as evidence of commercial trawl activity.
3) Collect a sediment sample at each station.
4) To collect samples of Nephrops for comparison of reproductive condition and morphometrics in each of the different survey areas (functional units). One trawl station per sediment stratum in each of the main survey areas.

This report focuses on the work up of the 2019 North Minch FU 11 UWTV survey. The abundance estimates are then used to provide catch options for the ICES Nephrops advice for 2020.

\section*{Methods}

The methods used in the survey are similar to those employed for other UWTV surveys of Nephrops stocks around Scotland and are documented by WKNEPHTV (ICES, 2007) and SGNEPS (ICES, 2010; ICES, 2012). A sledge is towed for a known distance and the number of burrows counted in a known field of view. Assuming a burrow occupancy of 1:1, the density of Nephrops is calculated and raised to the total habitat area.

Marine Scotland Science access to Vessel Monitoring System data (VMS) makes it possible to link geographical information on the positioning of vessels to landings data resulting in more detailed information on the spatial distribution of fishing effort in the Nephrops trawl fishery. For FU 11 the ground area calculation was based on the alpha convex-hull method to define and characterize the overall shape of a set of VMS points and is described in ICES (2010). The VMS area was updated in 2013 at the WKNEPH2013 (ICES, 2013a) and estimated to be \(2908 \mathrm{~km}^{2}\).

The survey design follows a random approach subject to certain geographic limits which ensure adequate coverage of the FU. At each sampling location the sledge is deployed and towed for 10 minutes. The distance covered by the sledge and the distance from the seabed are recorded to allow for the calculation of the viewed area. The video footage is recorded onto a DVD and analysed later by at least two counters providing independent estimates. Following the recommendation by SGNEPS (ICES, 2009), 7 minutes of footage are counted. All counters were re-familiarised using training material and reference footage for the North Minch before recounting at sea.

The Linn's concordance correlation coefficient (CCC) was used to compare the counts of the first two reviewers (obtained independently of each other). This statistical test measures correspondence between paired counts and has advantages over standard correlation analysis or \(t\) tests in that it measures agreement between two variables. For each run, a value of at least 0.5 correlation is expected for the first two counters and if this is not achieved the run is reviewed by more counters.

A number of factors are considered to influence the ability of the surveys to map directly to absolute abundance. In order to use the survey abundance estimate as an absolute estimate, it is necessary to apply a relative to absolute correction factor for these potential biases. The correction factors are FU specific and are given in the following table. They are based on simulation models, preliminary experimentation and expert opinion (ICES, 2009).
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline & Time period & Edge effect & detection rate & species identification & occupancy & Cumulative absolute factor & conversion \\
\hline FU 11:North Minch & Since 2009 & 1.38 & 0.85 & 1.1 & 1 & 1.33 & \\
\hline
\end{tabular}

All abundance estimates presented in this document are in absolute numbers (i.e. corrected for bias). Catch options for 2020 in FU 11 are provided using the 2019 abundance estimate.

\section*{Results}

A total of 47 valid survey stations were completed in the North Minch in 2019 out of 52. Two stations were aborted because of rocks and three stations were deemed to be off strata. Figure 1 shows the distribution of stations in the 2019 TV survey with the size of the symbols reflecting the Nephrops burrow density.

Table 1 gives the estimates of mean and variance for density and abundance for the most recent TV survey conducted in FU 11. From 2010 onwards, a single strata based on VMS was applied to calculate the overall abundance. The CV for the most recent TV survey is \(10.5 \%\) and is below the \(20 \%\) limit precision level recommended by SGNEPS.

Table 2 and Figure 2 show the time-series of the estimated abundance for the UWTV surveys, with \(95 \%\) confidence intervals on annual estimates. The abundance in 2019 ( 1232 million) shows an increase of \(3.7 \%\) compared to the value estimated for 2018 although the confidence intervals are larger.

Table 3 shows the assessment summary and table 4 the catch options inputs for 2020 with the most recent abundance estimate for FU 11. The forecast makes use of inputs derived from commercial data up to and including 2018. The inputs are the mean weight in landings and discards (1999-2018), the average discard rate (2016-2018) and the estimate of discard survival.

\section*{Discussion}

The UWTV survey is presented as the best available information on the North Minch Nephrops stock. The survey provides a fishery-independent estimate of Nephrops abundance. At present it is not possible to extract any length or age structure information from the survey and therefore it only provides information on abundance in numbers over the area of the survey.

The UWTV survey estimates of abundance for Nephrops in the North Minch suggest that historically the population increased until 2003. Over the next ten years it showed significant fluctuations. Since 2013, the abundance has been stable at around the average of the time series and was 1232 million in 2019.

The survey abundance shows fluctuations throughout the time series. While is not always possible to link these variations to fishing effort changes, other factors may contribute to this pattern such as variable recruitment being detected by the survey or variable natural mortality in this FU.

From 2016 the EU landing obligation was applied to all catches of Norway lobster fisheries in ICES Subarea 6, with several exemptions. There is a high survivability exemption applied to all Norway lobster creel fisheries and also for catches of Norway lobster made with demersal trawls using a cod end between \(80 \mathrm{~mm}-110 \mathrm{~mm}\) and within 12 miles of shore in ICES divisions 6a.
Observations from 2016-2018 fishery indicate that some discarding above the minimum conservation reference size (MCRS) continues and has not changed markedly. Consequently, ICES is providing advice for 2020 assuming average discard rates as observed over the three years, which is considered to be a more realistic assumption. ICES are also presenting advice assuming a zero discards scenario.

Two catch scenarios are presented in Table 5 and Table 6 . Table 5 assumes that selection parameters do not change. Table 6 assumes zero discarding. Under the landings obligation there are likely to be some changes in selectivity. However anecdotal information would indicate that the catch scenario assuming recent discard rates is most likely the best assumption. .

\section*{References}

ICES. 2007. Report of the Workshop on the use of UWTV surveys for determining abundance in Nephrops stocks throughout European waters (WKNEPHTV). ICES CM: 2007/ACFM: 14 Ref: LRC, PGCCDBS.

ICES. 2009. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 2009/LRC: 15, pp 52.

ICES. 2010. Report of the Study Group on Nephrops Surveys (SGNEPS), 9-11 November 2010, Lisbon, Portugal. ICES CM 2010/SSGESST:22. 95 pp.

ICES. 2012. Report of the Study Group on Nephrops Surveys (SGNEPS), 6-8 March 2012, Ancona, Italy. ICES CM 2012/SSGESST:19. 36 pp.

ICES. 2015. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE). ICES CM 2015/ACOM: 12

ICES. 2018. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 9-18 May 2018. ICES CM 2018/ACOM:13. 1340 pp.

ICES. 2019. Working Group for the Celtic Seas Ecoregion (WGCSE). ICES Scientific Reports. 1:29. 1078 pp. http://doi.org/10/17895/ices.pub. 7982.

2019


Figure 1. Nephrops, North Minch (FU11), TV survey station distribution and relative density (burrows/m²), 2019.

Catches


Figure 2. Nephrops, North Minch (FU11). Assessment summary.

Table 1. Nephrops, North Minch (FU 11): Results of the 2019 TV survey.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|}
\hline & & \(\stackrel{\square}{0}\) & \[
\begin{aligned}
& 3 \\
& 0 \\
& 0 \\
& \hline
\end{aligned}
\] & & & & \% & \\
\hline \[
\begin{aligned}
& E \\
& \text { E } \\
& \text { 䔍 }
\end{aligned}
\] & \[
\begin{aligned}
& \text { N } \\
& \stackrel{N}{\Sigma} \\
& \frac{1}{4} \\
& \hline
\end{aligned}
\] &  &  &  &  &  &  &  \\
\hline \multicolumn{9}{|l|}{2019 TV survey} \\
\hline VMS & 2908 & 47 & 0.423 & 0.092 & 1231.6 & 16438 & 1 & \\
\hline Total & 2908 & 47 & & & 1231.6 & 16438 & 1 & 0.10 \\
\hline
\end{tabular}

Table 2. Nephrops, North Minch (FU 11): Results of the UWTV surveys (absolute values).
\begin{tabular}{lllllll}
\hline & \begin{tabular}{l} 
Number \\
of valid
\end{tabular} & \begin{tabular}{l} 
Mean \\
density
\end{tabular} & \begin{tabular}{l} 
Abundance \\
(Sediment)
\end{tabular} & \begin{tabular}{l} 
95\% \\
confidence \\
interval
\end{tabular} & \begin{tabular}{l} 
Abundance \\
(VMS)
\end{tabular} & \begin{tabular}{l} 
Approx. \\
confidence \\
interval \\
(sediment)
\end{tabular} \\
& stations & & burrows/m
\end{tabular}

Table 3. Nephrops, North Minch (FU 11): Assessment summary.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline year & abundance millions & harvest.ratio \% & Landings numbers millions & Discard numbers millions & Removals numbers millions & landings tonnes & discard tonnes & Dead discard tonnes & Discard rate \% & mean.wt landings grammes & mean.wt discards grammes & dead.discard.rate \% \\
\hline 1994 & 820 & 31.5 & 154 & 139 & 258 & 3614 & 1637 & 1228 & 47.4 & 23.45 & 11.8 & 40.3 \\
\hline 1995 & NA & NA & 164 & 80 & 225 & 3655 & 856 & 642 & 32.8 & 22.24 & 10.65 & 26.8 \\
\hline 1996 & 541 & 23.5 & 108 & 26 & 127 & 2872 & 323 & 242 & 19.4 & 26.68 & 12.49 & 15.3 \\
\hline 1997 & NA & NA & 140 & 26 & 159 & 3046 & 286 & 215 & 15.4 & 21.71 & 11.18 & 12 \\
\hline 1998 & 898 & 12.2 & 103 & 8 & 110 & 2441 & 67 & 51 & 7.5 & 23.65 & 8.04 & 5.7 \\
\hline 1999 & 794 & 20.7 & 144 & 28 & 165 & 3257 & 273 & 205 & 16.4 & 22.7 & 9.69 & 12.8 \\
\hline 2000 & 1166 & 12.1 & 134 & 10 & 142 & 3247 & 100 & 75 & 6.9 & 24.19 & 10.08 & 5.2 \\
\hline 2001 & 1092 & 13 & 129 & 17 & 141 & 3259 & 160 & 120 & 11.7 & 25.33 & 9.32 & 9.1 \\
\hline 2002 & 1337 & 11.5 & 133 & 28 & 154 & 3440 & 277 & 208 & 17.6 & 25.93 & 9.78 & 13.8 \\
\hline 2003 & 1751 & 8.5 & 126 & 30 & 148 & 3269 & 299 & 224 & 19.2 & 26.03 & 10 & 15.2 \\
\hline 2004 & 1751 & 7.8 & 122 & 18 & 136 & 3082 & 202 & 151 & 13 & 25.16 & 11.02 & 10.1 \\
\hline 2005 & 1540 & 9.4 & 107 & 50 & 144 & 2949 & 507 & 380 & 32 & 27.65 & 10.09 & 26.1 \\
\hline 2006 & 1762 & 12.8 & 170 & 74 & 225 & 4166 & 757 & 568 & 30.3 & 24.52 & 10.27 & 24.6 \\
\hline 2007 & 1206 & 14.7 & 168 & 12 & 177 & 3978 & 214 & 160 & 6.5 & 23.61 & 18.1 & 5 \\
\hline 2008 & 1047 & 16.5 & 159 & 19 & 173 & 3799 & 194 & 145 & 10.5 & 23.9 & 10.36 & 8.1 \\
\hline 2009 & 1195 & 13.7 & 138 & 35 & 164 & 3496 & 327 & 245 & 20.3 & 25.42 & 9.34 & 16 \\
\hline 2010 & 1293 & 7 & 82 & 12 & 91 & 2413 & 128 & 96 & 12.4 & 29.39 & 10.98 & 9.6 \\
\hline 2011 & 1726 & 6.3 & 96 & 16 & 108 & 2697 & 154 & 116 & 14.2 & 27.56 & 9.66 & 11 \\
\hline 2012 & 891 & 18.7 & 151 & 21 & 167 & 3542 & 213 & 160 & 12.6 & 23.43 & 10.33 & 9.3 \\
\hline 2013 & 1403 & 10 & 122 & 24 & 140 & 3413 & 364 & 273 & 16.4 & 27.52 & 15.18 & 12.8 \\
\hline 2014 & 1251 & 9.6 & 115 & 8 & 121 & 3257 & 77 & 57 & 6.3 & 27.96 & 9.99 & 4.8 \\
\hline 2015 & 1445 & 7.9 & 103 & 15 & 114 & 3002 & 143 & 107 & 12.6 & 28.74 & 9.66 & 9.8 \\
\hline 2016 & 1422 & 10.7 & 136 & 22 & 152 & 3529.4 & 266 & 200 & 14 & 25.76 & 12.05 & 10.9 \\
\hline 2017 & 1050 & 9.3 & 93 & 5 & 97 & 2448 & 64 & 48 & 5.2 & 25.89 & 12.51 & 4 \\
\hline 2018 & 1188 & 6.4 & 72 & 5 & 76 & 1961 & 59 & 44 & 6.7 & 27.39 & 11.46 & 5.1 \\
\hline 2019 & 1232 & NA & NA & NA & NA & NA & NA & NA & NA & NA & NA & NA \\
\hline
\end{tabular}

Table 4. Nephrops, North Minch (FU 11): Basis for advice for 2020.
\begin{tabular}{|l|c|r|}
\hline \multicolumn{1}{|c|}{ Variable } & Value & Notes \\
\hline Stock abundance & 1232 million & Abundance in TV assessment UWTV 2019 \\
\hline Mean weight in wanted catch & 25.9 & Average 1999-2018 \\
\hline Mean weight in unwanted catch & 10.99 & Average 1999-2018 \\
\hline Unwanted catch proportion & \(8.6 \%\) & Average (proportion by number) 2016-2018 \\
\hline Unwanted catch survival rate & \(25 \%\) & Proportion by number \\
\hline Dead unwanted catch rate & \(6.6 \%\) & \begin{tabular}{r} 
Average 2016-2018 ( proportion by \\
number)
\end{tabular} \\
\hline
\end{tabular}

Table 5. Catch scenarios table for 2020 for FU 11 North Minch. Discarding assumed to continue at recent average. All weights in tonnes.
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline \multirow[b]{2}{*}{Basis} & Total catch & Dead removals & Wanted catch & \begin{tabular}{c} 
Dead \\
\begin{tabular}{c} 
unwanted \\
catch
\end{tabular} \\
\hline
\end{tabular} & Surviving unwanted catch & Harvest rate* & \multirow[b]{2}{*}{\% advice change **} \\
\hline & \[
\begin{gathered}
\text { WC+DUC } \\
+ \text { SUC }
\end{gathered}
\] & WC+DUC & WC & DUC & SUC & \[
\begin{gathered}
\text { for } \\
\text { WC+DU } \\
\text { C } \\
\hline
\end{gathered}
\] & \\
\hline \multicolumn{8}{|l|}{ICES advice basis} \\
\hline MSY approach & 3347 & 3315 & 3219 & 96 & 32 & 10.8 & 2.4\% \\
\hline \multicolumn{8}{|l|}{Other options} \\
\hline FMSY lower & 2604 & 2579 & 2504 & 75 & 25 & 8.4 & -20\% \\
\hline FMSY upper*** & 3347 & 3315 & 3219 & 96 & 32 & 10.8 & 2.4\% \\
\hline \(\mathrm{F}_{2018}\) & 1984 & 1965 & 1908 & 57 & 19 & 6.4 & -39\% \\
\hline
\end{tabular}
* Calculated for dead removals.
** Advice value 2020 relative to advice value 2019.
*** \(\mathrm{F}_{\text {MSY upper }}=\mathrm{F}_{\text {MSY }}\) for this stock

Table 6. Catch scenarios table for 2020 for FU 11 North Minch. Assuming zero discards. All weights in tonnes.
\begin{tabular}{|c|c|c|c|c|c|}
\hline \multirow[t]{2}{*}{Basis} & Total catch & Wanted catch & Unwanted catch & Harvest rate* & \multirow[t]{2}{*}{\begin{tabular}{l}
\% advice \\
change **
\end{tabular}} \\
\hline & WC+UC & WC & UC & for WC+UC & \\
\hline \multicolumn{6}{|l|}{ICES advice basis} \\
\hline MSY approach & 3276 & 3150 & 126 & 10.8 & 0.18\% \\
\hline \multicolumn{6}{|l|}{Other options} \\
\hline \(\mathrm{F}_{\text {MSY lower }}\) & 2548 & 2450 & 98 & 8.4 & -22\% \\
\hline FMSY upper*** & 3276 & 3150 & 126 & 10.8 & 0.18\% \\
\hline \(\mathrm{F}_{2018}\) & 1942 & 1867 & 75 & 6.4 & -41\% \\
\hline
\end{tabular}

\footnotetext{
* Calculated for dead removals.
** Advice value 2020 relative to advice value 2019.
*** FMSY upper \(=\) FMSY for this stock
}

\title{
Nephrops in the South Minch (FU12): results of the 2019 UWTV survey and catch options for 2020
}

\author{
Katie Boyle, Lynda Blackadder \& Helen Dobby \\ September 2019 \\ Marine Scotland Science \\ Aberdeen
}

\section*{Introduction}

Since the early 1990s Marine Scotland Science has carried out annual underwater television surveys (UWTV) targeting Nephrops norvegicus which have been used to estimate Nephrops abundance based on burrow counts. UWTV surveys for the estimation of Nephrops abundance reduce the problems associated with traditional trawl surveys that arise from variability in the burrow emergence of Nephrops. UWTV surveys are targeted at known areas of mud, sandy mud and muddy sand in which Nephrops construct burrows.

UWTV surveys provide an absolute estimate of stock abundance in numbers which is used to estimate the harvest ratio (a measure of exploitation rate), defined as the ratio of total catch to total abundance. Based on a harvest ratio equivalent to fishing at \(\mathrm{F}_{\text {MSY }}\) the landed proportion of the catch can be converted to biomass and is used by ICES to provide TAC advice.

UWTV surveys in the South Minch (FU 12) have been conducted since 1995. The survey takes place annually in May/June and also covers other Functional Units (FU) to the West of Scotland (FU 11 and FU 13) and in the North Sea (FU 7 and 34) and in 2019 FU10. The main objectives for the 2019 UWTV survey are listed below:
1) Obtain estimates of the abundance and distribution of Nephrops burrow complexes.
2) Use the video footage to record the occurrence of other benthic fauna as well as evidence of commercial trawl activity.
3) Collect a sediment sample at each station.
4) Collect samples of Nephrops for comparison of reproductive condition and morphometrics in each of the different survey areas (functional units). (One trawl station per sediment stratum in each of the main survey areas).

This report describes the work up of the 2019 South Minch (FU 12) UWTV survey. The abundance estimates are then used to provide catch options for ICES Nephrops advice for 2020.

\section*{Methods}

The methods used in the survey are similar to those employed for other UWTV surveys of Nephrops stocks around Scotland and are documented by WKNEPHTV (ICES, 2007) and SGNEPS (ICES, 2010; ICES, 2012). A sledge is towed for a known distance and the number of burrows counted in a known field of view. Assuming a burrow occupancy of 1:1, the density of Nephrops is calculated and raised to the total habitat area.

A random stratified sampling design is used. Stratification is on the basis of British Geological Survey (BGS) sediment strata - mud, sandy mud and muddy sand. The Nephrops sediment area in the South Minch is estimated to be \(5072 \mathrm{~km}^{2}\).

At each sampling location the sledge is deployed and towed for 10 minutes. The distance covered by the sledge and the distance from the seabed is recorded to allow for the calculation of the viewed area. The video footage is recorded onto a DVD and analysed later by at least two counters providing independent estimates. Following the recommendation by SGNEPS (ICES, 2009), 7 minutes of footage are counted. All counters were re-familiarised using training material and reference footage for the South Minch before recounting at sea.

The Linn's concordance correlation coefficient (CCC) was used to compare the counts (made independently) of the two reviewers. This statistical test measures correspondence between paired counts and has advantages over standard correlation analysis or \(t\) tests in that it measures agreement between two variables. For each run, a value of at least 0.5 correlation is expected for the first two counters and if this is not achieved the run is reviewed by more counters.

A number of factors are considered to influence the ability of the surveys to map directly to absolute abundance. In order to use the survey abundance estimate as an absolute value it is necessary to apply a relative to absolute correction factor to account for these potential biases. The correction factors are FU specific and for FU12 are given in the following table. They are based on a combination of the results of simulation models, preliminary experimentation and expert opinion (ICES, 2009).
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline & Time period & Edge effect & detection rate & species identification & occupancy & Cumulative absolute factor & conversion \\
\hline FU 12: South Minch & Since 2009 & 1.37 & 0.85 & 11 & & & \\
\hline
\end{tabular}

All abundance estimates presented in this document are in absolute numbers (i.e. corrected for bias). Catch options for 2020 in FU 12 are provided using the 2019 abundance estimate.

\section*{Results}

A total of 40 valid TV stations were completed in 2019, with one station excluded because of rocks. The number of stations in FU 12 UWTV survey has remained relatively stable throughout the time series. Figure 1 shows the distribution of stations in the 2019 TV survey with the size of the symbols reflecting the Nephrops burrow density. Higher densities were recorded in the inshore area of the ground, between the southwest of the Isle of Skye and the Ardnamurchan peninsula. Densities are generally lower in the western parts of the area towards the Outer Hebrides.

Table 1 gives the estimates of mean and variance for density and abundance (total and by strata) for the most recent UWTV survey conducted in FU 12. The CV for the most recent TV survey is \(12 \%\) and is below the \(20 \%\) limit precision level recommended by SGNEPS.

Table 2 and Figure 2 show the time-series of the estimated abundance for the UWTV surveys, with \(95 \%\) confidence intervals on annual estimates. The abundance in 2019 is 2362 million, which is an increase ( \(21 \%\) ) compared to the 2018 estimate.

Table 3 shows the assessment summary and table 4 the catch options inputs for 2020 with the most recent abundance estimate for FU 12. The forecast makes use inputs derived from commercial data up to and including 2018. The inputs are the mean weight in landings and discards (1999-2018), the average discard rate (2016-2018) and the estimate of discard survival.

\section*{Discussion}

The UWTV survey is presented as the best available information on the South Minch Nephrops stock. The surveys provide a fishery-independent estimate of Nephrops abundance. At present it is not possible to extract any length or age structure information from the survey and therefore it only provides information on abundance in numbers over the area of the survey.

The UWTV survey estimates of abundance for Nephrops in the South Minch show that the population has fluctuated without obvious trend over the period of the survey. The abundance decreased significantly in 2012 to the historical minimum but increased and was relatively stable at an above average level between 2014 and 2016. The abundance in 2019 ( 2362 million) shows a \(21 \%\) increase compared to the 2018 estimate.

The survey abundance has shown high fluctuations throughout the time series. While is not always possible to link this variations to fishing effort changes, other factors may contribute to this pattern such as variable recruitment being detected by the survey or variable natural mortality in this FU.

From 2016 the EU landing obligation was applied to all catches of Norway lobster fisheries in ICES Subarea 6, with several exemptions. There is a high survivability exemption applied to all Norway lobster creel fisheries and also for catches of

Norway lobster made with demersal trawls using a cod end between \(80 \mathrm{~mm}-110 \mathrm{~mm}\) and within 12 miles of shore in ICES divisions 6 a.
Observations from 2016-2018 fishery indicate that some discarding above the minimum conservation reference size (MCRS) continues and has not changed markedly. Consequently, ICES is providing advice for 2020 assuming average discard rates as observed over the three years, which is considered to be a more realistic assumption. ICES are also presenting advice assuming a zero discards scenario.

Two catch scenarios are presented in Table 5 and Table 6 . Table 5 assumes that selection parameters do not change. Table 6 assumes zero discarding. Under the landings obligation there are likely to be some changes in selectivity. However anecdotal information would indicate that the catch scenario assuming recent discard rates is most likely the best assumption.

\section*{References}

ICES. 2007. Report of the Workshop on the use of UWTV surveys for determining abundance in Nephrops stocks throughout European waters (WKNEPHTV). ICES CM: 2007/ACFM: 14 Ref: LRC, PGCCDBS.

ICES. 2009. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 2009/LRC: 15, pp 52.

ICES. 2010. Report of the Study Group on Nephrops Surveys (SGNEPS), 9-11 November 2010, Lisbon, Portugal. ICES CM 2010/SSGESST:22. 95 pp.

ICES. 2012. Report of the Study Group on Nephrops Surveys (SGNEPS), 6-8 March 2012, Acona, Italy. ICES CM 2012/SSGESST:19. 36 pp.

ICES. 2015. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE). ICES CM 2015/ACOM: 12

ICES. 2018. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE). 9-18 May 2018. ICES CM 2018/ACOM:13. 1340 pp.

ICES. 2019. Working Group for the Celtic Seas Ecoregion (WGCSE). ICES Scientific Reports. 1:29. 1078 pp. http://doi.org/10/17895/ices.pub. 7982.

\section*{2019}


Figure 1. Nephrops, South Minch (FU12), TV survey station distribution and relative density (burrows/m²), 2019. Crosses represent zero observations.


Figure 2. South Minch (FU12). Assessment summary.

Table 1. Nephrops South Minch (FU12): Results by stratum of the 2018 TV survey. Note that stratification was based on a series of sediment strata (M - Mud, SM - Sandy mud, MS - Muddy sand).


Table 2. Nephrops, South Minch (FU 12): Results of the UWTV surveys (absolute values).
\begin{tabular}{|c|c|c|c|c|}
\hline \multirow{4}{*}{Year} & \multirow{3}{*}{Stations} & \multicolumn{2}{|l|}{Mean} & \multirow[t]{2}{*}{\begin{tabular}{l}
95\% \\
confidence
\end{tabular}} \\
\hline & & density & Abundance & \\
\hline & & & & interval \\
\hline & & burrows/m \({ }^{2}\) & millions & millions \\
\hline 1995 & 33 & 0.23 & 1152 & 251 \\
\hline 1996 & 21 & 0.29 & 1473 & 530 \\
\hline 1997 & 36 & 0.21 & 1086 & 185 \\
\hline 1998 & 38 & 0.29 & 1452 & 232 \\
\hline 1999 & 37 & 0.21 & 1086 & 260 \\
\hline 2000 & 41 & 0.36 & 1854 & 348 \\
\hline 2001 & 47 & 0.4 & 2037 & 459 \\
\hline 2002 & 31 & 0.37 & 1899 & 567 \\
\hline 2003 & 25 & 0.42 & 2157 & 756 \\
\hline 2004 & 38 & 0.51 & 2558 & 473 \\
\hline 2005 & 33 & 0.43 & 2208 & 740 \\
\hline 2006 & 36 & 0.36 & 1845 & 598 \\
\hline 2007 & 39 & 0.2 & 1016 & 155 \\
\hline 2008 & 33 & 0.32 & 1608 & 415 \\
\hline 2009 & 25 & 0.3 & 1542 & 634 \\
\hline 2010 & 34 & 0.41 & 2076 & 665 \\
\hline 2011 & 36 & 0.38 & 1945 & 778 \\
\hline 2012 & 38 & 0.18 & 919 & 185 \\
\hline 2013 & 38 & 0.34 & 1718 & 365 \\
\hline 2014 & 36 & 0.41 & 2073 & 530 \\
\hline 2015 & 35 & 0.39 & 1998 & 514 \\
\hline 2016 & 37 & 0.42 & 2118 & 440 \\
\hline 2017 & 41 & 0.27 & 1384 & 282 \\
\hline 2018 & 39 & 0.38 & 1946 & 371 \\
\hline 2019 & 40 & 0.466 & 2362 & 578 \\
\hline
\end{tabular}

Table 3. Nephrops, South Minch (FU 12): Assessment summary.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline year &  &  &  &  &  &  &  &  & \[
\begin{aligned}
& \text { \# } \\
& \text { To } \\
& \text { O} \\
& \text { TH } \\
& 0 .
\end{aligned}
\] &  &  &  \\
\hline & millions & \% & millions & millions & millions & tonnes & tonnes & tonnes & \% & grammes & grammes & \% \\
\hline 1995 & 1152 & 20.9 & 213 & 37 & 241 & 4682 & 455 & 341 & 14.8 & 21.96 & 12.28 & 11.5 \\
\hline 1996 & 1473 & 14.2 & 173 & 48 & 209 & 3995 & 457 & 343 & 21.6 & 23.1 & 9.61 & 17.1 \\
\hline 1997 & 1086 & 19.3 & 186 & 31 & 209 & 4344 & 271 & 203 & 14.3 & 23.37 & 8.7 & 11.2 \\
\hline 1998 & 1452 & 13.3 & 168 & 32 & 192 & 3730 & 233 & 175 & 16.1 & 22.18 & 7.23 & 12.6 \\
\hline 1999 & 1086 & 16.9 & 161 & 29 & 183 & 4052 & 206 & 154 & 15.4 & 25.14 & 7 & 12 \\
\hline 2000 & 1854 & 9.2 & 145 & 33 & 170 & 3953 & 284 & 213 & 18.7 & 27.3 & 8.5 & 14.7 \\
\hline 2001 & 2037 & 10.6 & 168 & 65 & 216 & 3991 & 591 & 444 & 27.9 & 23.79 & 9.11 & 22.5 \\
\hline 2002 & 1899 & 7.5 & 123 & 26 & 143 & 3305 & 247 & 185 & 17.6 & 26.83 & 9.37 & 13.8 \\
\hline 2003 & 2157 & 7.8 & 139 & 38 & 168 & 3879 & 381 & 286 & 21.3 & 27.86 & 10.1 & 16.9 \\
\hline 2004 & 2558 & 6.8 & 141 & 44 & 175 & 3869 & 454 & 341 & 23.8 & 27.37 & 10.26 & 19 \\
\hline 2005 & 2208 & 7.9 & 137 & 49 & 174 & 3848 & 452 & 339 & 26.5 & 28.11 & 9.17 & 21.2 \\
\hline 2006 & 1845 & 10.8 & 177 & 30 & 199 & 4633 & 324 & 243 & 14.3 & 26.24 & 10.97 & 11.1 \\
\hline 2007 & 1016 & 27.3 & 228 & 66 & 278 & 5471 & 903 & 677 & 22.4 & 23.95 & 13.73 & 17.8 \\
\hline 2008 & 1608 & 17.4 & 224 & 74 & 279 & 5356 & 605 & 454 & 24.7 & 23.91 & 8.23 & 19.8 \\
\hline 2009 & 1542 & 12.9 & 179 & 26 & 199 & 4285 & 216 & 162 & 12.5 & 23.87 & 8.44 & 9.6 \\
\hline 2010 & 2076 & 7.6 & 149 & 12 & 158 & 3846 & 133 & 100 & 7.7 & 25.86 & 10.76 & 5.9 \\
\hline 2011 & 1945 & 6.5 & 118 & 11 & 126 & 3702 & 92 & 69 & 8.2 & 31.1 & 8.78 & 6.3 \\
\hline
\end{tabular}
\begin{tabular}{|cccccccccccccc|}
\hline \(\mathbf{2 0 1 2}\) & 919 & 15.8 & 133 & 16 & 145 & 3989 & 145 & 109 & 10.8 & 29.17 & 9.05 & 8.3 \\
\hline \(\mathbf{2 0 1 3}\) & 1718 & 8.1 & 136 & 4 & 140 & 3776 & 50 & 37 & 3.1 & 27.48 & 11.31 & 2.4 \\
\hline \(\mathbf{2 0 1 4}\) & 2073 & 5.8 & 105 & 19 & 120 & 3179 & 233 & 175 & 15.6 & 29.91 & 12.04 & 12.1 \\
\hline \(\mathbf{2 0 1 5}\) & 1998 & 6.4 & 120 & 10 & 128 & 3400 & 121 & 91 & 7.7 & 28.15 & 12.04 \\
\hline \(\mathbf{2 0 1 6}\) & 2118 & 9.5 & 177 & 31 & 201 & 4402 & 365 & 274 & 14.9 & 24.76 & 11.74 & 11.6 \\
\hline \(\mathbf{2 0 1 7}\) & 1384 & 9.9 & 127 & 13 & 137 & 3652 & 105 & 79 & 9.1 & 27.76 & 8.29 \\
\hline \(\mathbf{2 0 1 8}\) & 1946 & 4.8 & 91 & 4 & 94 & 2536 & 54 & 41 & 4.5 & 27.27 & 12.74 \\
\hline \(\mathbf{2 0 1 9}\) & 2362 & & & & & & & & & \\
\hline
\end{tabular}

Table 4. Nephrops, South Minch (FU 12): Basis for advice for 2020.
\begin{tabular}{|l|c|l|}
\hline \multicolumn{1}{|c|}{ Variable } & Value & \multicolumn{1}{c|}{ Notes } \\
\hline Stock abundance (2020) & 2362 million & Abundance in TV assessment UWTV 2019 \\
\hline \begin{tabular}{l} 
Mean weight in wanted \\
catch
\end{tabular} & 26.79 g & Average 1999-2018 \\
\hline \begin{tabular}{l} 
Mean weight in unwanted \\
catch
\end{tabular} & 10.08 g & Average 1999-2018 \\
\hline \begin{tabular}{l} 
Unwanted catch \\
proportion
\end{tabular} & \(9.5 \%\) & Average 2016-2018 (proportion by number). \\
\hline \begin{tabular}{l} 
Unwanted catch survival \\
rate
\end{tabular} & \(25 \%\) & Proportion by number. \\
\hline \begin{tabular}{l} 
Dead unwanted catch \\
rate*
\end{tabular} & \(7.3 \%\) & Average 2016-2018 (proportion by number). \\
\hline
\end{tabular}

Table 5. Catch scenarios table for 2019 for FU 12 South Minch. Discarding assumed to continue at recent average. All weights in tonnes.
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline \multirow[b]{2}{*}{Basis} & Total catch & Dead removals & Wanted catch & \[
\begin{gathered}
\text { Dead } \\
\text { unwanted } \\
\text { catch } \\
\hline
\end{gathered}
\] & Surviving unwanted catch & Harvest rate* & \multirow[b]{2}{*}{\% advice change **} \\
\hline & \[
\begin{gathered}
\text { WC+DUC } \\
+ \text { SUC }
\end{gathered}
\] & WC+DUC & WC & DUC & SUC & \[
\begin{gathered}
\text { for } \\
\text { WC+DU } \\
\text { C }
\end{gathered}
\] & \\
\hline \multicolumn{8}{|l|}{ICES advice basis} \\
\hline MSY approach & 7134 & 7066 & 6863 & 203 & 68 & 11.7 & 22\% \\
\hline \multicolumn{8}{|l|}{Other options} \\
\hline FMSY lower & 5671 & 5617 & 5455 & 162 & 54 & 9.3 & -3\% \\
\hline FMSY upper*** & 7134 & 7066 & 6863 & 203 & 68 & 11.7 & 22\% \\
\hline F2018 & 2927 & 2899 & 2816 & 83 & 28 & 4.8 & -50\% \\
\hline
\end{tabular}
* Calculated for dead removals.
** Advice value 2020 relative to advice value 2019.
*** \(\mathrm{F}_{\text {MSY upper }}=\mathrm{F}_{\text {MSY }}\) for this stock
All harvest rates are calculated in numbers and refer to the dead removals.

Table 6. Catch scenarios table for 2020 for FU 12 South Minch. Assuming zero discards. All weights in tonnes.
\begin{tabular}{|c|c|c|c|c|c|}
\hline \multirow[t]{2}{*}{Basis} & Total catch & Wanted catch & Unwanted catch & Harvest rate* & \multirow[t]{2}{*}{\% advice change **} \\
\hline & WC+UC & WC & UC & for WC+UC & \\
\hline \multicolumn{6}{|l|}{ICES advice basis} \\
\hline MSY approach & 6965 & 6700 & 265 & 11.7 & 19.2\% \\
\hline \multicolumn{6}{|l|}{Other options} \\
\hline \(\mathrm{F}_{\text {MSY lower }}\) & 5536 & 5326 & 210 & 9.3 & -22\% \\
\hline FMSY upper*** & 6965 & 6700 & 265 & 11.7 & -5.30\% \\
\hline F2018 & 2858 & 2749 & 109 & 4.8 & -51\% \\
\hline
\end{tabular}
* Calculated for dead removals.
** Advice value 2020 relative to advice value 2019.
*** \(\mathrm{F}_{\text {MSY upper }}=\mathrm{F}_{\text {MSY }}\) for this stock

\title{
Nephrops in the Clyde (FU13): results of the 2019 UWTV survey and catch options for 2020
}

\author{
Katie Boyle, Lynda Blackadder \& Helen Dobby \\ September 2019 \\ Marine Scotland Science \\ Aberdeen
}

\section*{Introduction}

Since the early 1990s Marine Scotland Science has carried out annual underwater television surveys (UWTV) targeting Nephrops norvegicus which have been used to estimate Nephrops abundance based on burrow counts. UWTV surveys for the estimation of Nephrops abundance reduce the problems associated with traditional trawl surveys that arise from variability in burrow emergence of Nephrops. UWTV surveys are targeted at known areas of mud, sandy mud and muddy sand in which Nephrops construct burrows.

UWTV surveys provide an absolute estimate of stock abundance in numbers which is used to estimate harvest ratio (a measure of exploitation rate), defined as the ratio of total catch in numbers by the total abundance. Based on a harvest ratio equivalent to fishing at \(\mathrm{F}_{\text {MSY }}\), the landed proportion of the catch can be converted to biomass and is used by ICES to provide TAC advice.

UWTV surveys in the Clyde FU 13 (Firth of Clyde and Sound of Jura) have been conducted for both sub areas since 1995 although the Sound of Jura has been surveyed more infrequently. The survey takes place annually in May/June and also covers other Functional Units (FU) to the west of Scotland (FU 11 and FU 12) and in the North Sea (FU 7 and 34) and in 2019 FU10. The main objectives for the 2019 UWTV survey are listed below:
1) Obtain estimates of the abundance and distribution of Nephrops burrow complexes.
2) Use the video footage to record the occurrence of other benthic fauna as well as evidence of commercial trawl activity.
3) Collect a sediment sample at each station.
4) Collect samples of Nephrops for comparison of reproductive condition and morphometrics in each of the different survey areas (functional units). (One trawl station per sediment stratum in each of the main survey areas).

This report documents the work up of the 2019 FU 13 UWTV survey which includes two subareas: the Firth of Clyde and the Sound of Jura. The abundance estimates of the two sub areas are worked up separately and used in separate catch options tables for the ICES Nephrops advice for 2020.

\section*{Methods}

The methods used in the survey are similar to those employed for other UWTV surveys of Nephrops stocks around Scotland and are documented by WKNEPHTV (ICES, 2007) and SGNEPS (ICES, 2010; ICES, 2012). A sledge is towed for a known distance and the number of burrows counted in a known field of view. Assuming a 1:1 burrow occupancy, the density of Nephrops is calculated and raised to the total habitat area.

A random stratified sampling design is used for both subareas. Stratification is on the basis of British Geological Survey (BGS) sediment strata - mud, sandy mud and muddy sand. The Nephrops sediment area in FU 13 is estimated to be \(2081 \mathrm{~km}^{2}\) for the Firth of Clyde and \(382 \mathrm{~km}^{2}\) for the Sound of Jura.

At each sampling location the sledge is deployed and towed for 10 minutes. The distance covered by the sledge and the distance from the seabed are recorded to allow for the calculation of the viewed area. The video footage is recorded onto a DVD and analysed later by at least two counters providing independent estimates. Following the recommendation by SGNEPS (ICES, 2009), 7 minutes of footage are counted. All counters were re-familiarised using training material and reference footage for the Clyde and Sound of Jura before recounting at sea.

The Linn's concordance correlation coefficient (CCC) was used to compare the counts (made independently) of the two reviewers. This statistical test measures correspondence between paired counts and has advantages over standard correlation analysis or \(t\) tests in that it measures agreement between two variables. For each run, a value of at least 0.5 correlation is expected for the first two counters and if this is not achieved the run is reviewed by more counters

A number of factors are considered to influence the ability of the surveys to map directly to absolute abundance. In order to use the survey abundance estimate as an absolute estimate, it is necessary to apply a relative to absolute correction factor for these potential biases. The correction factors are FU specific and are given in the following table. They are based on simulation models, preliminary experimentation and expert opinion (ICES, 2009).
\begin{tabular}{llllll|l} 
& \multicolumn{4}{c}{ Time period Edge effect } & detection rate & \begin{tabular}{l} 
species \\
identification
\end{tabular}
\end{tabular} occupancy \begin{tabular}{l} 
Cumulative \\
absolute \\
factor
\end{tabular}\(\quad\) conversion

All abundance estimates presented in this document are in absolute numbers (i.e. corrected for bias). Catch options for 2020 in FU 13 are provided using the 2019 abundance estimate.

\section*{Results}

The total number of valid TV stations completed was 38 for the Firth of Clyde and 12 for the Sound of Jura (similar numbers to previous years). Figure 1 shows the distribution of stations for the 2019 UWTV surveys in FU 13, with the size of the symbols reflecting the Nephrops burrow density.

Tables 1 and 2 give the estimates of mean and variance for density and abundance (total and by strata) for the Firth of Clyde and Sound of Jura respectively. The CVs for the most recent TV survey in the Firth of Clyde and Sound of Jura are within the \(20 \%\) limit precision level proposed by SGNEPS at \(9 \%\) and \(10 \%\) respectively.

Tables 3 and 4 and Figure 2 show the time-series of the estimated abundances for both subareas, with \(95 \%\) confidence intervals on annual estimates. The abundance in 2019 in the Firth of Clyde has decreased by around \(5 \%\) since 2018 and is now 2083 million. For the Sound of Jura, the abundance estimate is 318 million which is an increase of \(15 \%\) since 2018.

Table 5 shows the assessment summary and table 6 the catch options inputs for 2020 with the most recent abundance estimates for each of the subareas in FU 13. The historical harvest rates were calculated using catches and abundance for the whole FU combined. This combined harvest rate is considered to be more representative for the Firth of Clyde than the Sound of Jura.

The forecast makes use inputs derived from commercial data up to and including 2018. The inputs are the mean weight in landings and discards (1999-2018), the average discard rate (2016-2018) and the estimate of discard survival.

\section*{Discussion}

The UWTV survey is presented as the best available information on the FU 13 Nephrops stocks. The surveys provide a fishery-independent estimate of Nephrops abundance. At present it is not possible to extract any length or age structure information from the survey and therefore it only provides information on abundance over the area of the survey.

The UWTV survey estimates of abundance for Nephrops in the Firth of Clyde suggest that the population increased until the mid 2000s suggesting a sustained period of increased recruitment. Over the past 10 years, the abundance has fluctuated around the long term average value. The 2019 abundance estimate for the Firth of Clyde is 2083 million.

Historically in the Sound of Jura, the abundance has fluctuated without trend. However, since 2013 it showed a continual increase to one of the highest values in the time series in 2016, before falling in 2017 and 2018 with an increase in 2019 to 318 million.

Since 2015, the assessment has provided an estimate of the two subareas combined, rather than separately. This is because it is not possible to reliably disaggregate the landings (and catch) data for the two sub-areas. In contrast, catch options are provided for each subarea separately.

From 2016 the EU landing obligation was applied to all catches of Norway lobster fisheries in ICES Subarea 6, with several exemptions. There is a high survivability exemption applied to all Norway lobster creel fisheries and also for catches of Norway lobster made with demersal trawls using a cod end between \(80 \mathrm{~mm}-110 \mathrm{~mm}\) and within 12 miles of shore in ICES divisions 6a.
Observations from 2016-2018 fishery indicate that some discarding above the minimum conservation reference size (MCRS) continues and has not changed markedly. Consequently, ICES is providing advice for 2020 assuming average discard rates as observed over the three years, which is considered to be a more realistic assumption. ICES are also presenting advice assuming a zero discards scenario.

Two catch scenarios are presented in Table 5 and Table 6 . Table 5 assumes that selection parameters do not change. Table 6 assumes zero discarding. Under the landings obligation there are likely to be some changes in selectivity. However anecdotal information would indicate that the catch scenario assuming recent discard rates is most likely the best assumption.

\section*{References}

ICES. 2007. Report of the Workshop on the use of UWTV surveys for determining abundance in Nephrops stocks throughout European waters (WKNEPHTV). ICES CM: 2007/ACFM: 14 Ref: LRC, PGCCDBS.

ICES. 2009. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 2009/LRC: 15, pp 52.

ICES. 2010. Report of the Study Group on Nephrops Surveys (SGNEPS), 9-11 November 2010, Lisbon, Portugal. ICES CM 2010/SSGESST:22. 95 pp.

ICES. 2012. Report of the Study Group on Nephrops Surveys (SGNEPS), 6-8 March 2012, Acona, Italy. ICES CM 2012/SSGESST:19. 36 pp.

ICES. 2015. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE). ICES CM 2015/ACOM: 12

ICES. 2018. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE). 9-18 May 2018. ICES CM 2018/ACOM:13. 1340 pp.

ICES. 2019. Working Group for the Celtic Seas Ecoregion (WGCSE). ICES Scientific Reports. 1:29. 1078 pp. http://doi.org/10/17895/ices.pub. 7982.

\section*{2019}


Figure 1. Nephrops, Clyde (FU13), TV survey station distribution and relative density (burrows \(/ \mathrm{m}^{2}\) ) for Firth of Clyde and Sound of Jura subareas, 2019. Sound of Jura located in the west side. Shaded green and brown areas represent areas of suitable sediment for Nephrops.


Figure 2. Clyde (FU13): Long term trends in catch (tonnes), harvest rate \& UWTV survey abundance (by sub-area separately).

Table 1. Nephrops, Clyde (FU 13): Firth of Clyde subarea. Results by sediment stratum for 2019 TV survey. Note that stratification was based on a series of sediment strata (M - Mud, SM - Sandy mud, MS - Muddy sand).


Table 2. Nephrops, Clyde (FU 13): Sound of Jura subarea. Results by sediment stratum for 2018 TV survey. Note that stratification was based on a series of sediment strata (M - Mud, SM - Sandy mud, MS - Muddy sand).


Table 3. Nephrops, Clyde (FU 13): Firth of Clyde subarea. Results of the UWTV surveys (absolute values).


Table 4. Nephrops, Clyde (FU 13): Sound of Jura subarea. Results of the UWTV surveys (absolute values).
\begin{tabular}{|c|c|c|c|c|}
\hline \multirow{4}{*}{Year} & \multicolumn{3}{|c|}{Mean} & 95\% \\
\hline & \multirow[t]{3}{*}{Stations} & \multirow[t]{2}{*}{density} & \multirow[t]{2}{*}{Abundance} & confidence \\
\hline & & & & interval \\
\hline & & burrows/m² & millions & millions \\
\hline 1995 & 7 & 0.42 & 160 & 58 \\
\hline 1996 & 10 & 0.45 & 171 & 26 \\
\hline \multicolumn{5}{|l|}{1997} \\
\hline 1998 & & & & \\
\hline \multicolumn{5}{|l|}{1999 (} \\
\hline \multicolumn{5}{|l|}{2000} \\
\hline 2001 & 13 & 0.71 & 272 & 76 \\
\hline 2002 & 9 & 1.04 & 398 & 167 \\
\hline 2003 & 12 & 0.68 & 260 & 68 \\
\hline 2004 & \multicolumn{4}{|l|}{no survey} \\
\hline 2005 & 11 & 0.79 & 303 & 84 \\
\hline 2006 & 10 & 1.13 & 430 & 134 \\
\hline 2007 & 10 & 0.67 & 255 & 58 \\
\hline 2008 & \multicolumn{4}{|l|}{no survey} \\
\hline 2009 & 12 & 0.66 & 251 & 68 \\
\hline 2010 & 12 & 0.98 & 376 & 38 \\
\hline 2011 & 12 & 0.82 & 312 & 73 \\
\hline 2012 & 12 & 0.98 & 371 & 61 \\
\hline 2013 & 9 & 0.52 & 198 & 35 \\
\hline 2014 & 9 & 0.61 & 231 & 90 \\
\hline 2015 & 12 & 0.98 & 376 & 127 \\
\hline 2016 & 12 & 1.10 & 422 & 42 \\
\hline 2017 & 12 & 0.80 & 306 & 71 \\
\hline 2018 & 12 & 0.72 & 275 & 53 \\
\hline 2019 & 12 & 0.832 & 318 & 61 \\
\hline
\end{tabular}

Table 5. Nephrops, Clyde (FU13) (Firth of Clyde and Sound of Jura subareas); Assessment summary
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline \[
\begin{aligned}
& \text { п } \\
& \stackrel{\pi}{x}
\end{aligned}
\] &  &  &  &  &  &  &  &  &  &  & \[
\begin{aligned}
& \text { 은 } \\
& \stackrel{H}{0} \\
& \stackrel{H}{0}
\end{aligned}
\] &  &  &  \\
\hline & millions & & millions & & \% & millions & millions & millions & tonnes & tonnes & \% & grammes & grammes & \% \\
\hline 1995 & 579 & 176 & 160 & 58 & 36.4 & 207 & 82 & 269 & 3987 & 619 & 28.4 & 19.24 & 7.54 & 22.9 \\
\hline 1996 & 935 & 242 & 171 & 26 & 21.1 & 187 & 61 & 233 & 4057 & 635 & 24.7 & 21.68 & 10.35 & 19.7 \\
\hline 1997 & 1198 & 262 & NA & NA & NA & 150 & 70 & 202 & 3621 & 598 & 32 & 24.21 & 8.5 & 26.1 \\
\hline 1998 & 1262 & 213 & NA & NA & NA & 269 & 187 & 409 & 4841 & 1292 & 41 & 17.98 & 6.92 & 34.2 \\
\hline 1999 & 930 & 289 & NA & NA & NA & 216 & 93 & 286 & 3752 & 566 & 30.2 & 17.39 & 6.05 & 24.5 \\
\hline 2000 & 1411 & 246 & NA & NA & NA & 171 & 48 & 207 & 3417 & 470 & 22 & 19.96 & 9.75 & 17.4 \\
\hline 2001 & 1486 & 268 & 272 & 76 & 12.8 & 164 & 82 & 225 & 3182 & 677 & 33.5 & 19.46 & 8.23 & 27.4 \\
\hline 2002 & 1571 & 288 & 398 & 167 & 12.4 & 207 & 50 & 245 & 3384 & 406 & 19.5 & 16.35 & 8.12 & 15.4 \\
\hline 2003 & 1817 & 292 & 260 & 68 & 12.8 & 166 & 134 & 266 & 3173 & 1247 & 44.7 & 19.13 & 9.31 & 37.7 \\
\hline 2004 & 1970 & 367 & NA & NA & NA & 158 & 168 & 284 & 2973 & 1435 & 51.5 & 18.8 & 8.54 & 44.3 \\
\hline 2005 & 1959 & 287 & 303 & 84 & 10.7 & 189 & 69 & 241 & 3395 & 611 & 26.8 & 17.96 & 8.81 & 21.6 \\
\hline 2006 & 1851 & 257 & 430 & 134 & 12.7 & 248 & 55 & 290 & 4780 & 515 & 18.2 & 19.27 & 9.31 & 14.3 \\
\hline 2007 & 1233 & 218 & 255 & 58 & 43 & 350 & 387 & 640 & 6660 & 2566 & 52.5 & 19.05 & 6.64 & 45.3 \\
\hline 2008 & 1769 & 291 & NA & NA & NA & 357 & 207 & 512 & 5923 & 1433 & 36.6 & 16.59 & 6.94 & 30.3 \\
\hline 2009 & 1499 & 210 & 251 & 68 & 22.2 & 261 & 169 & 388 & 4779 & 1390 & 39.3 & 18.31 & 8.23 & 32.7 \\
\hline 2010 & 1750 & 327 & 376 & 38 & 14.9 & 276 & 55 & 317 & 5843 & 536 & 16.7 & 21.21 & 9.68 & 13.1 \\
\hline 2011 & 2165 & 305 & 312 & 73 & 15.7 & 333 & 74 & 388 & 6432 & 568 & 18.2 & 19.34 & 7.65 & 14.3 \\
\hline 2012 & 1421 & 227 & 371 & 61 & 21 & 306 & 93 & 376 & 6687 & 1066 & 23.4 & 21.83 & 11.42 & 18.6 \\
\hline 2013 & 1990 & 246 & 198 & 35 & 14.1 & 262 & 62 & 309 & 5435 & 454 & 19 & 20.72 & 7.37 & 15 \\
\hline
\end{tabular}
\begin{tabular}{|cccccccccccccc|}
\(\mathbf{2 0 1 4}\) & 1328 & 237 & 231 & 90 & 22.6 & 295 & 78 & 353 & 6207 & 696 & 20.9 & 20.79 & 8.92 \\
\hline \(\mathbf{2 0 1 5}\) & 1820 & 351 & 376 & 127 & 12.4 & 232 & 54 & 273 & 5147 & 401 & 18.9 & 22.21 & 7.43 \\
\hline \(\mathbf{2 0 1 6}\) & 1946 & 249 & 422 & 42 & 17.6 & 364 & 69 & 416 & 6447 & 636 & 15.9 & 17.7 & 9.21 \\
\(\mathbf{2 0 1 7}\) & 1568 & 239 & 306 & 71 & 17.6 & 305 & 31 & 329 & 5222 & 265 & 9.2 & 17.0 & 8.55 \\
\(\mathbf{2 0 1 8}\) & 2193 & 297 & 275 & 53 & 11.1 & 268 & 7 & 273 & 4141 & 68 & 2.5 & 16.14 & 9.79 \\
\(\mathbf{2 0 1 9}\) & 2083 & 381 & 318 & 61 & NA & NA & NA & NA & NA & NA & NA & NA & NA \\
\hline
\end{tabular}

Table 6. . Nephrops, Clyde (FU 13): Basis for advice for 2020.
Firth of Clyde
\begin{tabular}{|l|r|l|}
\hline \multicolumn{1}{|c|}{ Variable } & \multicolumn{1}{|c|}{ Value } & \multicolumn{1}{|c|}{ Notes } \\
\hline Stock abundance (2020) & 2083 million & Abundance in TV assessment UWTV 2019 \\
\hline Mean weight in wanted catch & 16.95 g & \begin{tabular}{l} 
Average 2016-2018 (combined for Firth of Clyde \\
and Sound of Jura)
\end{tabular} \\
\hline Mean weight in unwanted catch & 9.18 g & \begin{tabular}{l} 
Average 2016-2018 (combined for Firth of Clyde \\
and Sound of Jura)
\end{tabular} \\
\hline Unwanted catch proportion & \(9.2 \%\) & \begin{tabular}{l} 
Average proportion by number 2016-2018 \\
(combined for Firth of Clyde and Sound of Jura)
\end{tabular} \\
\hline Unwanted catch survival rate & \(25 \%\) & Proportion by number. \\
\hline Dead unwanted catch rate* & \(7.1 \%\) & Average 2016-2018 (proportion by number). \\
\hline
\end{tabular}

Sound of Jura
\begin{tabular}{|l|r|l|}
\hline \multicolumn{1}{|c|}{ Variable } & \multicolumn{1}{|c|}{ Value } & \multicolumn{1}{c|}{ Notes } \\
\hline Stock abundance (2020) & 318 million & Abundance in TV assessment UWTV 2019 \\
\hline Mean weight in wanted catch & 16.95 g & \begin{tabular}{l} 
Average 2016-2018 (combined for Firth of Clyde \\
and Sound of Jura)
\end{tabular} \\
\hline Mean weight in unwanted catch & 9.18 g & \begin{tabular}{l} 
Average 2016-2018 (combined for Firth of Clyde \\
and Sound of Jura)
\end{tabular} \\
\hline Unwanted catch proportion & \(9.2 \%\) & \begin{tabular}{l} 
Average proportion by number 2016-2018 \\
(combined for Firth of Clyde and Sound of Jura)
\end{tabular} \\
\hline Unwanted catch survival rate & \(25 \%\) & Proportion by number. \\
\hline Dead unwanted catch rate* & \(7.1 \%\) & Average 2016-2018 (proportion by number). \\
\hline
\end{tabular}

Table 7. Catch options tables for 2019 for Firth of Clyde subarea. All weights in tonnes.
Catch scenarios for 2020 assuming discarding continues at the recent average rate.
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline \multirow[b]{2}{*}{Basis} & Total catch & Dead removals & Wanted catch & \[
\begin{gathered}
\text { Dead } \\
\text { unwanted } \\
\text { catch } \\
\hline
\end{gathered}
\] & Surviving unwanted catch & Harvest rate* & \multirow[b]{2}{*}{\% advice change **} \\
\hline & \[
\begin{gathered}
\text { WC+DUC } \\
+ \text { SUC }
\end{gathered}
\] & WC+DUC & WC & DUC & SUC & \[
\begin{gathered}
\text { for } \\
\text { WC+DU } \\
\text { C } \\
\hline
\end{gathered}
\] & \\
\hline \multicolumn{8}{|l|}{ICES advice basis} \\
\hline MSY approach & 5227 & 5159 & 4955 & 204 & 68 & 15.1 & -12.70\% \\
\hline \multicolumn{8}{|l|}{Other options} \\
\hline FMSY lower & 3428 & 3383 & 3249 & 134 & 45 & 9.9 & -43\% \\
\hline \(\mathrm{F}_{\text {MSY }}\) upper*** & 5227 & 5159 & 4955 & 204 & 68 & 15.1 & -12.70\% \\
\hline \(\mathrm{F}_{2018}\) & 3842 & 3792 & 3642 & 150 & 50 & 11.1 & -36\% \\
\hline
\end{tabular}
* Calculated for dead removals and applied to total catch.
**Advice value 2019 relative to advice value 2018.
*** \(\mathrm{F}_{\text {msY upper }}=\mathrm{F}_{\text {msy }}\) for this stock
Catch scenarios for 2020 assuming zero discards.
\begin{tabular}{|c|c|c|c|c|c|}
\hline Basis & Total catch
WC+UC & Wanted catch
WC & Unwanted
catch & Harvest rate*
for \(\mathrm{WC}+\mathrm{UC}\) & \% advice change ** \\
\hline \multicolumn{6}{|l|}{ICES advice basis} \\
\hline MSY approach & 5107 & 4841 & 266 & 15.1 & -14.70\% \\
\hline \multicolumn{6}{|l|}{Other options} \\
\hline FMSY lower & 3348 & 3174 & 174 & 9.9 & -44\% \\
\hline F MSY upper*** & 5107 & 4841 & 266 & 15.1 & -14.70\% \\
\hline \(\mathrm{F}_{2018}\) & 3754 & 3559 & 195 & 11.1 & -37\% \\
\hline
\end{tabular}
* Calculated for dead removals.
** Advice value 2020 relative to advice value 2019.
*** \(\mathrm{F}_{\text {MSY upper }}=\mathrm{F}_{\text {MSY }}\) for this stock

Table 8. Catch options tables for 2020 for Sound of Jura subarea. All weights in tonnes.
Catch scenarios for 2020 assuming discarding continues at the recent average rate.
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline \multirow[b]{2}{*}{Basis} & Total catch & Dead removals & Wanted catch & \[
\begin{gathered}
\text { Dead } \\
\text { unwanted } \\
\text { catch }
\end{gathered}
\] & Surviving unwanted catch & Harvest rate* & \multirow[b]{2}{*}{\(\%\) advice change **} \\
\hline & \[
\begin{aligned}
& \text { WC+DUC } \\
& \text { +SUC }
\end{aligned}
\] & WC+DUC & WC & DUC & SUC & \[
\begin{gathered}
\text { for } \\
\text { WC+DU } \\
\text { C }
\end{gathered}
\] & \\
\hline \multicolumn{8}{|l|}{ICES advice basis} \\
\hline MSY approach & 634 & 626 & 601 & 25 & 8 & 12 & 6.00\% \\
\hline \multicolumn{8}{|l|}{Other options} \\
\hline FMSY lower & 496 & 490 & 471 & 19 & 6 & 9.4 & -17.10\% \\
\hline FMSY upper*** & 634 & 626 & 601 & 25 & 8 & 12 & 6.00\% \\
\hline \(\mathrm{F}_{2018}\) & 587 & 579 & 556 & 23 & 8 & 11.1 & -1.84\% \\
\hline
\end{tabular}
* Calculated for dead removals and applied to total catch.
**Advice value 2019 relative to advice value 2018.
*** \(\mathrm{F}_{\text {MSY upper }}=\mathrm{F}_{\text {MSY }}\) for this stock
Catch scenarios for 2020 assuming zero discards.
\begin{tabular}{|c|c|c|c|c|c|}
\hline \multirow[t]{2}{*}{Basis} & Total catch & Wanted catch & Unwanted
catch & Harvest rate* & \multirow[t]{2}{*}{\% advice change **} \\
\hline & WC+UC & WC & UC & for WC+UC & \\
\hline \multicolumn{6}{|l|}{ICES advice basis} \\
\hline MSY approach & 619 & 587 & 32 & 12.0 & 3.50\% \\
\hline \multicolumn{6}{|l|}{Other options} \\
\hline FmsY lower & 485 & 460 & 25 & 9.4 & -18.90\% \\
\hline FMSY upper*** & 619 & 587 & 32 & 12.0 & 3.50\% \\
\hline \(\mathrm{F}_{2018}\) & 573 & 543 & 30 & 11.1 & -4.2\% \\
\hline
\end{tabular}
* Calculated for dead removals.
** Advice value 2020 relative to advice value 2019.
*** \(\mathrm{F}_{\text {MSY upper }}\) = FMSy for this stock


\title{
Western Irish Sea Nephrops Grounds (FU15) 2019 UWTV Survey Report and catch options for 2020
}

Mathieu Lundy \({ }^{1}\), Peter McCorriston \({ }^{1}\), Ian McCausland \({ }^{1}\), Keith Erskine \({ }^{1}\), Katie Lilley \({ }^{1}\), Gary Heaney \({ }^{1}\),Jim McArdle \({ }^{1}\), Aaron Buick \({ }^{1}\), Jessica Graham \({ }^{1}\), Charlotte Reeve \(^{3}\) and Jennifer Doyle \({ }^{2}\)
\({ }^{1}\) Fisheries and Aquatic Ecosystems Branch, Agri-Food \& Biosciences Institute, Newforge Lane, Belfast BT9 5PX, Northern Ireland.
\({ }^{2}\) Fisheries Ecosystems Advisory Services, The Marine Institute, Renville, Oranmore, Galway, Ireland.
\({ }^{3}\) Centre for Environment, Fisheries and Aquaculture Science, Pakefield Road, Lowestoft, Suffolk NR33 0HT, England.


\begin{abstract}
This report provides the main results and findings of the \(17^{\text {th }}\) annual underwater television survey on the 'Irish sea west Nephrops grounds' ICES assessment area, Functional Unit 15. The survey was multi-disciplinary in nature collecting UWTV and other ecosystem data. The 2019 design consisted of a randomised isometric grid of 100 stations at 4.5 nautical mile intervals out over the full known extent the stock. The resulting krigged burrow abundance estimate was 4.4 billion burrows. This was a similar result of that obtained in 2015, but a \(10 \%\) lower than the abundance in 2018. In contrast to 2017 the spatial distribution of burrows shows a high density band on the central western area of the survey ground. The abundance remains within previously observed ranges and is above MSY \(\mathrm{B}_{\text {trigger. }}\) The CV (or relative standard error) of \(3 \%\) is in line with previous estimates and well below the upper limit of \(20 \%\) recommended by SGNEPS 2012. Total catches and landings options at various different fishing mortalities were calculated and fishing at \(\mathrm{F}_{\text {msy }}\) in 2020 implies a total catch option at \(\mathrm{F}_{\text {msy }}\left(=\mathrm{F}_{\max }\right)\) of 10,377 tonnes estimated to result in landings of no more than 8,546 tonnes. Sea-pens were observed at \(21 \%\) of stations with high densities observed in the south-west of the ground. Trawl marks were noted at \(15 \%\) of the UWTV stations.
\end{abstract}

Key words: Nephrops norvegicus, stock assessment, geostatistics, underwater television (UWTV), benthos.

\section*{Introduction}

The Norway lobster, Nephrops norvegicus, is exploited throughout its geographic range, from Icelandic waters to the Mediterranean and the Moroccan coast. The western Irish Sea stock (FU15) is amongst the most productive of all the Nephrops stocks currently fished yielding landings of \(5,000-10,000\) tonnes annually from a relatively small geographic area (ICES, 2012a). Nephrops spend significant time in burrows, with emergence behaviour influenced by several factors: time of year, light intensity, tidal strength, etc. Underwater towed video surveys and assessment methodologies have been developed to provide a fishery independent estimate of stock size, exploitation status and catch advice (ICES, 2009a \& 2012a).

This is the \(17^{\text {th }}\) survey in a time series of UWTV surveys in the western Irish Sea carried out jointly by the Agri-Food and Biosciences Institute (AFBI), Northern Ireland, and the Marine Institute, Ireland. The 2019 survey was multi-disciplinary in nature; the specific objectives are listed below:
1. To complete randomised fixed isometric survey grid of 100 UWTV with 4.5 nautical mile (nm) spacing stations on the western Irish Sea Nephrops ground (FU15).
2. To obtain 2019 quality assured estimates of Nephrops burrow distribution and abundance on the western Irish Sea Nephrops ground (FU15). These will be compared with those collected previously.
3. To collect ancillary information from the UWTV footage at each station such as the occurrence of sea-pens, other macro-benthos and fish species and trawl marks on the seabed.
4. Technology, staff and protocol transfer between AFBI, the Marine Institute and Cefas.

This report details the final UWTV results of the 2019 survey and also documents other data collected during the survey.

SGNEPS (ICES, 2012b) recommended that a CV (or relative standard error) of \(<20 \%\) is an acceptable precision level for UWTV surveys. SGNEPS also recommended that investigations into the precision of surveys be carried out and where possible survey effort should be extended to grounds not already covered with UWTV surveys (including FU16, FU19 and FU20-21).

\section*{Material and methods}

From 2003 to 2019 a randomized fixed square grid for the western Irish Sea (FU15) Nephrops ground has been used. An adaptive approach is taken whereby stations are continued past the known perimeter of the ground until the burrow densities are zero or very close to zero. The initial ground perimeter has been established using a combination of integrated logbook-VMS data (using the methods described in Gerritsen and Lordan, 2011), British Geological Survey (BGS) and other sediment maps, and previously collected UWTV data. The same ground boundaries have been used throughout the time-series. The grid spacing from 2003 to 2011 was 3.5 nautical miles (nM). Following a review (Doyle et al., 2013) the grid design was changed from a 3.5 nM to 4.5 nM in 2012. In 2013, the grid spacing was increased further to a 5.0 nM
isometric grid, whereas a 4.5 nM isometric grid was used again in 2014-2018 to ensure all edge of ground areas were represented adequately.

The survey took place on RV Corystes between \(6^{\text {th }}\) August and \(16^{\text {th }}\) August 2019. The survey covered the western Irish Sea (FU15) grid and the eastern Irish Sea (FU14). The results for FU14 will be presented in a separate report, led by Cefas. Survey timing for FU15 was generally standardised to August each year and was also timed to take full advantage of the neap tides when underwater visibility is normally better.

The protocols used were those presented and reviewed by WKNEPHTV 2007 (ICES, 2007) and are summarised as follows: at each station the UWTV sledge was deployed and once stable on the seabed a 10 minute tow was recorded onto DVD. Vessel position (dGPS) were recorded every 1 second. The navigational data were quality controlled using an " \(R\) " script developed by the Marine Institute (ICES, 2009b). In 2019 due to technical issues with the USBL, ship GPS navigational data were used to calculate distance over ground for \(100 \%\) of stations, this was as carried out in 2018.

In line with SGNEPS recommendations all scientists were trained/re-familiarised using training material and validated using reference footage prior to recounting at sea (ICES, 2009b). Once this process had been undertaken, all recounts were conducted by two trained "burrow identifying" scientists independent of each other on board the research vessel during the survey. During this review process the visibility, ground type and speed of the sledge during one-minute intervals were subjectively classified using a classification key. In addition the numbers of Nephrops burrows complexes (multiple burrows in close proximity which appear to be part of a single complex which are only counted once), Nephrops activity in and out of burrows were counted and recorded by each scientist for each one-minute interval. Following the SISP (Series of ICES Survey Protocols) agreed by WGNEPS 2016 (ICES, 2016), eight minutes of recounts should be carried out for each station although only the first minute is then excluded from analysis with only seven minutes used for analysis. The first minute is thus treated as a 're-familiarisation' minute.

Notes were also recorded each minute on the occurrence of trawl marks, fish species and other species. Semi-quantitative abundance of seapen species were also recorded according to OSPAR Special Request (ICES 2011). A key was devised to categorise the densities of sea pens based on SACFOR abundance scale (Table 2) after ICES (2011). Finally, if there was any time during the one-minute where counting was not possible (due to sediment clouds or other reasons), the duration was also recorded so that the time window could be removed from the distance over ground calculations. The " R " quality control tool allowed for individual station data to be analysed in terms of data quality for navigation, overall tow factors such as speed and visual clarity and consistency in counts (examples are given in Figures 1 and 2). Consistency and bias between individual counters were examined no obvious bias between counters was observed.

The recount data were subjected to Lin's concordance correlation coefficient (CCC) for each station, and where the statistic fell below a threshold of 0.5 , a third independent trained counter completed an additional set of recounts. These recounts were checked again using Lin's CCC and if the statistic remained below the threshold of 0.5 , a fourth consensus count was carried out by the three counters who had previously counted the
footage. This application of Lin's CCC to recount data during the survey is the fourth year where this has been used to bring quality control in line with other laboratories' protocols, such as those of Marine Scotland, Marine Institute and Cefas. The "R" scripts were developed by AFBI based on those used by Cefas. Lin's CCC was only applied to stations where average count per minute exceeded 1.5 burrow systems. In total, 35 of the stations failed to meet the 0.5 Lin's CCC threshold based on their first two counts by independent counters. Following third counts over \(96 \%\) stations successfully met these threshold, however 4 stations did not pass after having 3 counts by independent staff and a consensus count was carried out for each of these stations.

Arithmetic means of the burrow density and Nephrops recounts were standardised by dividing by the survey area observed. The ship's navigation data were used to calculate distance over ground of the sledge. The field of view of the camera at the bottom of the screen was estimated at 68 cm using lasers with the sledge flat on the seabed (i.e. no sinking).

The " R " Geostats was used to complete the interpolation and analysis of data, and to calculate the CV. This was completed as in previous years using core " R " code used by the Marine Institute and first trialled on FU15 in 2016, using the adjusted burrow densities. A historical comparison of SURFER based estimates and R-geostats estimates showed consistent trend agreement. The Coefficient of Variance is provided using R-Geostats for the 2019 dataset.

\section*{Results}

The station positions are shown in Figure 3. A violin plot of the observed burrow densities from 2003 to 2019 on the western Irish Sea is presented in Figure 4. Over the time series available the density estimates observed are very similar with average density of around 1 burrow \(/ \mathrm{m}^{2}\). Figure 5 and Figure 6 show the variability in density between minutes and operators (counters) for each station, respectively. These quality control and consistency plots show that the burrow estimates were fairly consistent between minutes and counters. Variability is higher between minutes than counters. Higher density stations showed the greatest variability between counters. Stations in the west and south-west of the ground tend to show higher minute by minute variability than those in the centre and to the north of the ground. Recent trawling activity and the co-occurrence of other burrowing species (e.g. Goneplax rhomboides and Calocaris macandreae) sometimes impacts on the between minute variability.

The geo-statistical structural analysis is shown in the form of variograms in Figure 9. The blanked and krigged contour plot and posted point density data are shown in Figures \(8-10\). The krigged contours correspond well to the observed data. These densities surfaces show a relatively dynamic situation. Some parts of the ground have consistently higher or lower densities, such as to the south-west of the ground, near the northern-most extent of the ground, with a further 'hot spot' to the east of the ground (southwest of the Isle of Man). In most areas densities drop to zero or near zero as the ground boundary is approached, with the exception in 2016-2018 (and to a lesser extent in 2014) across the widest part of the ground at the western and easten (immediately SW of Isle of Man) boundaries. There tends to be a lower density towards the centre of the ground. The 2019 spatial pattern is most similar to that in 2007, but with slighty lower densities observed towards the southern west of the ground. The high density areas observed in the FU15 in the past had almost disappeared in recent years, however 2016 and 2017 has exhibited some stations with an average burrow density of greater than 2 systems \(/ \mathrm{m}^{2}\); but in 2019 there were no densities this high observed, with the densities similar to those observed in 2015 (see Figure 5).

The summary statistics from this geo-statistical analysis are given in Table 3 and plotted in Figure 11. The 2019 final adjusted abundance estimate of 4.4 billion burrows is very close to that estimated in 2015, but represent some of the lowest total abundances on record. The overall burrow abundance trend is fairly stable although the abundance did decline between 2005 and 2008, and some decline was also seen between 2012 and 2015 but increased in 2016, and again in 2017, before falling in 2018. The CV for 2018 was \(3 \%\) indicating a very precise survey in line with CVs observed previously.

Seapen distribution across the western Irish Sea Nephrops grounds is mapped in Figure 12. Trawl marks were noted at \(26 \%\) of the stations surveyed, which is a decrease of \(10 \%\) to that observed in the previous survey year.

\section*{Discussion}

The western Irish Sea (FU 15) stock has accounted for \(>40 \%\) of the total landings reported to WGCSE for ICES Sub-area VII (ICES, 2018) making it an important FU in the TAC management area. The burrow densities typically observed in FU15 are amongst the highest observed of all Nephrops stocks but the mean sizes of individuals
in the catches are relatively small. It appears that growth is suppressed due to competition and/or recruitment effects (Johnson et. al, 2012). There has been an increase in the mean landed weight of individual Neprhops, although a complex interaction of environmental and fishery factors could cause this. Despite the smaller size of individuals, the fishery is particularly important to the Northern Irish and Irish Nephrops métiers. In the last decade it has become by far the most economically important fishery in the Irish Sea. The Western Irish Sea Nephrops stock is relatively well studied with size information on catches extending back to the 1970s, a trawl survey series since 1994 and larval production surveys in a few years.

Since the benchmark assessment by ICES in 2009 this UWTV survey has become the main input for assessment and calculation of catch options for this stock. The survey information up to 2012 was used as the main basis for the ICES assessment of status and exploitation rate up to 2012. ICES concluded that this stock abundance is stable and is above MSY \(\mathrm{B}_{\text {trigger }}\) (ICES, 2013). The 2019 abundance estimate remains well above the MSY \(\mathrm{B}_{\text {trigger }}\) (biomass trigger) proposed by ICES of 3.0 billion burrows which was derived from a longer time series of trawl survey data. All other stock status indicators suggest that the stock remains at a stable healthy condition (ICES, 2013). Table 4 is an updated management option table giving total catch and landings options at various levels of fishing mortality for 2020. Using the 2019 estimate of abundance would imply a total catch option at \(\mathrm{F}_{\text {MSY }}\left(=\mathrm{F}_{\max }\right)\) of no more 10,377 tonnes which would result in landings of more than 8,546 tonnes, if the fisher behaviour (discarding) as observed in 2016-2018 was maintained.

SGNEPS 2012 recommended a review of survey sampling intensity (ICES, 2012b). Following a review (Doyle et al., 2013) the grid design was changed from a 3.5 nautical mile square grid prior to 2012 to 5.0 nautical mile isometric grid in 2013. In 2014 the grid spacing was reduced to 4.5 nautical mile isometric grid as in 2012, and has remained at this spacing in the following years. The precision for all surveys at 4.5 nM spacing appears stable and high, with a CV of \(3 \%\) which is in line with previous estimates well below the SGNEPS 2012 recommendation of \(20 \%\).

Burrow identification in the western Irish Sea is, at times, difficult due to the high underlying burrow densities and sometimes poor visibility. The burrows of Calocaris macandreae (a mud burrowing shrimp species) are abundant particularly in the softer muds in the middle of the western Irish Sea grounds, and the burrows of the crab Goneplax rhomboides in the west and south-west in particular cause some difficulty with Nephrops burrow discrimination. However, such allocation errors are minimised due to the training procedures employed during the survey. These include refresher training on classical Nephrops burrow signatures and consistency verification with reference count analyses (ICES, 2008 \& 2009b). The counting performance of the 2018 counters was generally very high with Lin's CCC scores \(>0.5\) for all stations.

An important objective of this UWTV survey is to collect various ancillary information. The occurrence of trawl marks on the footage is notable for two reasons. Firstly, it makes identification of Nephrops burrows more difficult as the trawl marks can remove some signature features making accurate burrow identification more difficult. Secondly, only occupied Nephrops burrows will persist in heavily trawled grounds and it is assumed that each burrow is occupied by one individual Nephrops (ICES, 2009a). The CTD data collected is currently being analysed as part of a larger project. The
multi-disciplinary nature of the survey means that the information collected is highly relevant for a number of research and advisory applications.

The impact of trawling activity on the seabed communities' structure and functioning has been raised a potential ecosystem concern (OSPAR, 2010). Seapens in particular have been identified as a potential indicator species for benthic habitat health status. OSPAR have sought advice from ICES on the utility of UWTV surveys for collecting data on seapen status and distributions (ICES, 2011). The occurrence of seapens has been noted on this survey since the outset. This is the sixth year that a systematic quantification and of seapens was undertaken. There is evidence of co-occurrence of trawl marks and seapens, particularly in the south of the ground.

The main objectives of the survey were successfully met for the 17 th successive year. The UWTV coverage and footage quality were generally good on the western Irish Sea grounds due to survey timing. Due to fishing activity over some stations these sites had to be re-visited as first attempts were marred by poor visibility. There were some technical difficulties with camera equipment.

\section*{Acknowledgments}

We would like to express our thanks and gratitude to the Captain and crew of RV Corystes for their good will and professionalism during the survey. Finally, thanks to the AFBI, Marine Institute, and CEFAS staff onboard for their hard work and enthusiasm in making this survey a success.

\section*{References}

Doyle, J., Lordan, C., Fitzgerald, R., Strong, J. and Service, M. 2012. Western Irish Sea Nephrops Grounds (FU15) 2012 UWTV Survey Report. Marine Institute and Agri-Food and Biosciences Institute. http://hdl.handle.net/10793/891
Doyle J., Lordan C., Ligas A., Brown V., Leocadio A., McCausland I., McCorriston P., Service M., Stewart P., and Schön P.-J. 2013. Western Irish Sea Nephrops Grounds (FU15) 2013 UWTV Survey Report and catch options for 2014. Marine Institute and AFBI UWTV Survey report.
Gerritsen, H., and Lordan, C. 2011. Integrating vessel monitoring systems (VMS) data with daily catch data from logbooks to explore the spatial distribution of catch and effort at high resolution. ICES Journal of Marine Science: Journal du Conseil 68: 245-252. doi: http://dx.doi.org/10.1093/icesjms/fsq137.
Petitgas P. and Lafont, T, 1997. EVA (Estimation VAriance). A geostatistical software on IBM-PC for structure characterization and variance computation. Version 2.
ICES 2007. Report of the Workshop on the use of UWTV surveys for determining abundance in Nephrops stocks throughout European waters (WKNEPHTV). ICES CM: 2007/ACFM: 14 Ref: LRC, PGCCDBS.
ICES 2008. Report of the Workshop and training course on Nephrops burrow identification (WKNEPHBID). ICES CM: 2008/LRC: 3 Ref: LRC, ACOM.
ICES 2009a. Report of the Benchmark Workshop on Nephrops assessment (WKNEPH). ICES CM: 2009/ACOM:33

ICES 2009b. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 2009/LRC: 15. Ref: TGISUR.
ICES 2011. Report of the ICES Advisory Committee 2011. ICES Advice.2011. Book 1: Introduction, Overviews and Special Requests. Protocols for assessing the status of sea-pen and burrowing megafauna communities, section 1.5.5.3.
ICES 2012a. Report of the Working Group for Celtic Seas Ecoregion (WGCSE). ICES CM: 2012/ ACOM:12.
ICES 2012b. Report of the Study Group on Nephrops Surveys (SGNEPS). ICES CM 2012/SSGESST: 19. Ref: SCICOM, ACOM
ICES 2012c. ACOM advice June 2012, Nephrops in Subarea VII, ICES Advice 2012, Book 5, pp 234-283.
ICES 2013. Report of the Working Group for Celtic Seas Ecoregion (WGCSE). ICES CM: 2013/ ACOM:12.
ICES 2016. Interim Report of the Working Group on Nephrops Surveys (WGNEPS). ICES CM 2016/SSGIEOM:32.
Johnson, M.P., Lordan, C., Power. A.M. 2012. Habitat and Ecology of Nephrops norvegicus. Advances in marine biology 64, 27-63
OSPAR, 2010. Background Document for Seapen and Burrowing megafauna communities. OSPAR Commission 2010, London. Publication number: 481/2010


Figure 1: FU15 western Irish Sea grounds: R - tool quality control plot for station 34 of the 2018 survey.


Figure 2: FU15 western Irish Sea grounds: R - tool quality control plot for station 26 of the 2019 survey.


Figure 3: FU15 western Irish Sea grounds: Stations completed on the 2019 UWTV Survey.


Figure 4: FU15 western Irish Sea grounds: Violin plot of burrow density distributions by year from 2003-2018.


Figure 5: FU15 western Irish Sea grounds: Plot of the variability in density between minutes for each station in 2018.


Figure 6: FU15 western Irish Sea grounds: Plot of the variability in density between operators (counters) for each station in 2018.

\section*{2019 FU15 UWTV Density auto fit Exponential}


Figure 7: FU15 western Irish Sea grounds: Variograms for 2019.


Figure 8: FU15 western Irish Sea grounds: Contour plots of the krigged density estimates by year from 2003-2008.


Figure 9: FU15 western Irish Sea grounds: Contour plots of the krigged density estimates by year from 2009-2014.


Figure 10: FU15 western Irish Sea grounds: Contour plots of the krigged density estimates by year from 2015-2018.


Figure 11: FU15 western Irish Sea grounds: Time series of geo-statistical adjusted abundance estimates (in billions of burrows) from 2003-2019. Error bars correspond to the \(95 \%\) confidence intervals calculated in EVA. Blue horizontal line is \(B_{\text {trigger }}\) of 3.0 billion burrows.


Figure 13: FU15 western Irish Sea grounds: Stations where sea pens were present in was identified during 2019.

Table 1: Key for classification of Seapen abundance as used on Irish UWTV surveys.
Number/Min
Common 20-200
Frequent 2-19
Ocasional <2

Species
Virgularia mirabilis
Pennatula phosphorea
Funiculina quadrangularis


Table 2: Cumulative bias factors for each Nephrops stock surveyed by UWTV method.
\begin{tabular}{|l|c|ccc|c|c|}
\cline { 2 - 7 } \multicolumn{1}{c|}{} & & & & & \\
\multicolumn{1}{c|}{} & Edge effect & \begin{tabular}{c} 
Burrow \\
detection
\end{tabular} & \begin{tabular}{c} 
Burrow \\
identification
\end{tabular} & \begin{tabular}{c} 
Burrow \\
occupancy
\end{tabular} & FU & \begin{tabular}{c} 
Cumulative \\
Bias
\end{tabular} \\
\hline 3\&4 Skagerrak and Kattegat (IIIa) & 1.3 & 0.75 & 1.05 & 1 & FU3 & 1.1 \\
6:Farn Deeps & 1.3 & 0.85 & 1.05 & 1 & FU6 & 1.2 \\
7:Fladen & 1.45 & 0.9 & 1 & 1 & FU7 & 1.35 \\
8:Firth of Forth & 1.23 & 0.9 & 1.05 & 1 & FU8 & 1.18 \\
9:Moray Firth & 1.31 & 0.9 & 1 & 1 & FU9 & 1.21 \\
10: Noup & 1.31 & 0.9 & 1 & 1 & FU10 & 1.21 \\
11:North Minch & 1.38 & 0.85 & 1.1 & 1 & FU11 & 1.33 \\
12:South Minch & 1.37 & 0.85 & 1.1 & 1 & FU12 & 1.32 \\
13:Clyde & 1.19 & 0.75 & 1.25 & 1 & FU13 & 1.19 \\
14: Irish Sea East & 1.3 & 0.85 & 1.05 & 1 & FU14 & 1.2 \\
15:Irish Sea West & 1.24 & 0.75 & 1.15 & 1 & FU15 & 1.14 \\
16: Porcupine & 1.26 & 0.95 & 1.05 & 1 & FU16 & 1.26 \\
17:Aran & 1.35 & 0.9 & 1.05 & 1 & FU17 & 1.3 \\
19:South Coast & 1.25 & 0.9 & 1.15 & 1 & FU19 & 1.3 \\
20\&21 Labadie & 1.25 & 0.9 & 1.15 & 1 & FU20 & 1.3 \\
22:Smalls & 1.35 & 0.9 & 1.05 & 1 & FU22 & 1.3 \\
34: Devil's Hole & 1.3 & 0.85 & 1.05 & 1 & FU34 & 1.2 \\
\hline
\end{tabular}

Table 3: FU15 western Irish Sea grounds: Overview of geo-statistical results from 2003-2019.
\begin{tabular}{|c|c|c|c|c|c|c|}
\hline FU 15 & Year & Numbe \(r\) of station s & Mean Density adjusted (burrows./m²) & Domain Area ( \(\mathrm{km}^{2}\) ) & \(\stackrel{\text { R- }}{\text { GeoStats }}\) Estimate & CV on Burrow estimate \\
\hline \multirow[t]{17}{*}{\[
\begin{gathered}
\text { Western Irish } \\
\text { Sea }
\end{gathered}
\]} & 2003 & 160 & 0.99 & 5295 & 5.8 & 3\% \\
\hline & 2004 & 147 & 1.00 & 5310 & 5.8 & 3\% \\
\hline & 2005 & 141 & 1.02 & 5281 & 6.1 & 4\% \\
\hline & 2006 & 138 & 0.97 & 5194 & 5.9 & 4\% \\
\hline & 2007 & 148 & 0.93 & 5285 & 5.4 & 3\% \\
\hline & 2008 & 141 & 0.77 & 5287 & 4.5 & 3\% \\
\hline & 2009 & 142 & 0.83 & 5267 & 4.9 & 3\% \\
\hline & 2010 & 149 & 0.90 & 5307 & 5.1 & 3\% \\
\hline & 2011 & 156 & 0.88 & 5289 & 5.2 & 2\% \\
\hline & *2012 & 99 & 0.91 & 5291 & 5.5 & 3\% \\
\hline & *2013 & 80 & 0.78 & 5278 & 4.5 & 3\% \\
\hline & *2014 & 99 & 0.83 & 5272 & 4.6 & 3\% \\
\hline & *2015 & 100 & 0.79 & 5279 & 4.4 & 3\% \\
\hline & *2016 & 100 & 0.84 & 5260 & 5.2 & 3\% \\
\hline & *2017 & 101 & 0.90 & 5304 & 5.4 & 3\% \\
\hline & *2018 & 100 & 0.85 & 5791 & 4.9 & 3\% \\
\hline & *2019 & 100 & 0.73 & 5792 & 4.4 & 3\% \\
\hline
\end{tabular}

Table 4: FU15 western Irish Sea grounds: Short-term forecast management option table giving catch options for 2018 using the 2017 UWTV survey estimate.
\begin{tabular}{|l|c|c|c|c|c|}
\hline \multirow{2}{*}{ Basis } & \begin{tabular}{c} 
Total \\
catches*
\end{tabular} & Landings & \begin{tabular}{c} 
Dead \\
discards**
\end{tabular} & \begin{tabular}{c} 
Surviving \\
discards**
\end{tabular} & \begin{tabular}{c} 
Harvest \\
rate
\end{tabular} \\
\cline { 2 - 6 } & L+DD+SD & L & \multicolumn{1}{c|}{ DD } & SD & for L+DD \\
\hline F \(_{\text {MSY }}\) & 10,377 & 8,546 & 1,648 & 0,183 & 18.2 \\
\hline \(\mathrm{~F}_{2018}\) & 5,701 & 4,696 & 905 & 0,101 & 10.0 \\
\hline FmsY lower \(^{\text {FmSY upper precautionary }}\) & 10,377 & 8,546 & 1,122 & 0,125 & 12.4 \\
\hline
\end{tabular}

\footnotetext{
Weights in tonnes.
* Total catches are the landings plus dead and surviving discards.
** Total discard rate is assumed to be \(28.5 \%\) of the catches (in number); discard survival is assumed to be \(10 \%\).
}

\title{
01_2019_FU2021_SISP__datacheck \\ MAE/JD \\ 2019-09-23
}

This markdown documment compares the original sea count data for 2019 FU2021 and 20\% SISP check stations counted after the survey.
- This review was undertaken as a result of plenary discussions at WGNEPS 2018 in May and WGNEPS 2018 SISP review, where it was recommended by the WG to review \(20 \%\) of stations when there is a substantial difference in abundance from the previous year.
- The \(20 \%\) review stations were distributed randomly among the 4 person review team.
- All reviews were undertaken in the lab after there was some initial training and familiaristion.

I am using R version 3.6.1 (2019-07-05). First we load the required packages.
```

library(RODBC)
library(sqldf)
library(epiR)
library(reshape2)
library(mapplots)
library(shapefiles)
library(gridExtra)
library(grid)
library(lme4)
library(tidyverse)
library(vioplot)
library(data.table)
library(maptools)
library(maps)
library(mapproj)
library(captioner)
library(knitr)

```

Now to check the versions of the different packages:
\begin{tabular}{ll}
\hline Package & Version \\
\hline RODBC & \(1.3-15\) \\
sqldf & \(0.4-11\) \\
epiR & \(1.0-2\) \\
reshape2 & 1.4 .3 \\
mapplots & 1.5 .1 \\
shapefiles & 0.7 \\
gridExtra & 2.3 \\
grid & 3.6 .1 \\
lme4 & \(1.1-21\) \\
tidyverse & 1.2 .1 \\
vioplot & 0.3 .2 \\
data.table & 1.12 .2 \\
maptools & \(0.9-5\) \\
maps & 3.3 .0
\end{tabular}
\begin{tabular}{ll}
\hline Package & Version \\
\hline mapproj & 1.2 .6 \\
captioner & 2.2 .3 \\
knitr & 1.23 \\
\hline
\end{tabular}

Set Up Tables and Figures

\section*{Sample random stations}

A random selection of stations from 2019 FU2021 footage where this selection does not include any true zero stations, that is, non nephrops habitat such as gravels, shelly sand etc. In total, we will review 19 stations from 2019 FU2021 footage.

Station 240 and 259 were chcosen as training stations - where all 4 reviewers counted these and discussed each burrow.
```

setwd("N:/Surveys/UWTV SURVEYS FU2022 CELTIC SEA/2019 Labadie/SISP_qc_check")
real.stn <- c(173,176,181,183,194,198,204,209,210,216,223,225,232,240,253,256,258,259,1850)
knitr::kable(cbind(Year=rep("2019", 19), Station=real.stn))

```
\begin{tabular}{ll}
\hline Year & Station \\
\hline 2019 & 173 \\
2019 & 176 \\
2019 & 181 \\
2019 & 183 \\
2019 & 194 \\
2019 & 198 \\
2019 & 204 \\
2019 & 209 \\
2019 & 210 \\
2019 & 216 \\
2019 & 223 \\
2019 & 225 \\
2019 & 232 \\
2019 & 240 \\
2019 & 253 \\
2019 & 256 \\
2019 & 258 \\
2019 & 259 \\
2019 & 1850 \\
\hline
\end{tabular}

\section*{Map showing the 2019 stations and the \(20 \%\) SISP check stations.}
```


## Read the data and subset the year of interest

dat <- fread("N:/Surveys/UWTV SURVEYS FU2022 CELTIC SEA/2019 Labadie/Kriging/fu2021_tv_final_2019.cSv",
dat<- subset(dat, Year==2019)
FG <- readShapePoly('//Galwayfs03/Nephrops/Surveys/ArcGis/ShapefilesR/NephropsGrounds_All',proj4string=

## Warning: readShapePoly is deprecated; use rgdal::readOGR or sf::st_read

EU <- readShapePoly('//Galwayfs03/Nephrops/Surveys/ArcGis/ShapefilesR/Europe',proj4string=CRS('+proj=lo

## Warning: readShapePoly is deprecated; use rgdal::readOGR or sf::st_read

m <- ggplot() +
geom_polygon(data=EU, aes(x=long, y=lat, group=group), fill = "\#006837") +
geom_polygon(data=FG, aes(x=long, y=lat, group=group),fill = "Light Grey")
latlimits <- c(49.6, 51.2)
longlimits <- c(-7, -9.2)
dat <- dat %>% rename(lon=MidDeglong) %>% rename(lat=MidDegLat)

# new dataframe

dat2 <- dat
dat2<-data.frame(dat2)
str(dat2)

## 'data.frame': 95 obs. of 8 variables:

## \$ FU : int 2021 2021 2021 2021 2021 2021 2021 2021 2021 2021 ...

## \$ Survey : chr "CV19022" "CV19022" "CV19022" "CV19022" ...

## \$ Year : int 2019 2019 2019 2019 2019 2019 2019 2019 2019 2019 ...

## \$ Station: int 170 171 172 173 174 175 176 177 178 179 ...

## \$ Count : num 0 0 16.7 5.5 0 ...

## \$ lon : num -8.13 -7.98 -8.36 -8.21 -8.05 ...

## \$ lat : num 49.7 49.7 49.8 49.8 49.8 ...

## \$ Ground : chr "Labadie" "Labadie" "Labadie" "Labadie" ...

class(dat2)

## [1] "data.frame"

dat3 <- subset(dat2, Station %in% real.stn)
m +
\#geom_point(data=dat2 , aes(x=lon, y=lat), shape =3, colour = "blue") +

```
```

    geom_point(data=dat3, aes(x=lon, y=lat), shape=1, size=8, colour="red") +
    geom_text(data = dat2, aes(x=lon, y=lat, label = (Station)), size = 4) +
    theme_bw() +
coord_cartesian(xlim = longlimits, ylim = latlimits) +
labs(y="Latitude",x="Longitude")

```


Figure 1: Map showing 2019 stations and historical review stations denoted by red circle

\section*{Read in the validation data}
```

setwd("N:/Surveys/UWTV SURVEYS FU2022 CELTIC SEA/2019 Labadie/SISP_qc_check")
\#ensure delete additional minutes in local survey database
\#run queries 1-5 to generate Recounts-Clean or pop them in here to code
channel <- odbcConnectAccess("N:/Surveys/UWTV SURVEYS FU2022 CELTIC SEA/2019 Labadie/SISP_qc_check/LABA
dataframe<-sqlFetch(channel, "Recounts")
recounts.clean <- dataframe[-which (dataframe$StartTime - dataframe$StopTime > 30),]
stns <- read.csv("SISP_Labadie_2019_stations_assigned_to_counters.csv")
stns <- stns[1:19,c("stn","first","second")]
stns19 <- stns
rec <- recounts.clean[which(recounts.clean$Video_Line_Name %in% unique(stns19$stn)),]
Recounts <- subset(rec, VideoOperatorID %in% c(18,19,20,21))
Ops<-sqlFetch(channel, "Video_Operator")

```

\section*{Check number of recount stations}
```

length(unique(Recounts\$Video_Line_Name))

## [1] 19

Recounts <- Recounts[order(Recounts$Video_Line_Name, Recounts$SurveyID,Recounts$VideoOperatorID,Recount
selst <- Recounts %>% group_by(Video_Line_Name) %>%
    summarise(hatn =sum(BurrowCount)/length(BurrowCount)) %>%
    filter(hatn> 1.5)
selst <- as.list(selst[1])
SummedBurrow<-aggregate(BurrowCount~Video_Line_Name, sum, data = Recounts)
ZeroBurrow<-subset(SummedBurrow, BurrowCount > 0)
RecountsP<-subset(Recounts, Video_Line_Name %in% selst$Video_Line_Name)
length(unique(Recounts\$Video_Line_Name))

## [1] 19

```

\section*{Quality control using Lin's CCC test.}

As standard on UWTV surveys run the counts to check counter performance using a threshold of 0.5 for FU2021.
```

Lins<- 0.5
Oid <- sort(unique(RecountsP$Video_Line_Name))
par(mfrow=c (3,2))
for (o in c(1:length(Oid)))
{
    temp<-RecountsP[RecountsP$Video_Line_Name==Oid[o],]
temp <- temp[order(temp$Minute),]
    Rs <- unique(temp$VideoOperatorID)
l <- length(Rs)
if(l>1)
{
for( i in c(1:(l-1)))
{
for(j in c((i+1):l))
{
temp2 <- temp[temp$VideoOperatorID %in% c(Rs[i],Rs[j]),]
                temp2<-dcast(temp2,Minute~VideoOperatorID, value.var = "BurrowCount", sum)
            c1<-as.numeric(names(temp2[2]))
            c2<-as.numeric(names(temp2[3]))
            c1<-Ops$Initials[match(c1,0ps$VideoOperatorID)]
            c2<-Ops$Initials[match(c2,Ops$VideoOperatorID)]
            tmp.ccc <- epi.ccc(temp2[,2], temp2[,3], ci = "z-transform",conf.level = 0.95)
            z <- lm(temp2[,3] ~ temp2[,2])
            par(pty = "s")
            plot(temp2[,2],temp2[,3], xlab = c1,ylab = c2, pch = 16,
                    main = paste("Video Line =", unique(temp$Video_Line_Name),
"; Lin's CCC =",round(tmp.ccc$rho.c[1],2), sep = " "))
            abline(a = 0, b = 1, lty = 2)
            abline(z, lty = 1)
            text(temp2[,2]+.1,(temp2[,3]),temp2[,1], cex = 0.6)
            plot(temp2$Minute, temp2[,2], type="b", col=1, pch=0, lty=1,xlab="Minute",
ylab = "Burrows counted", xlim=range(temp2$Minute), ylim=range(temp2[,2:3]))
            points(temp2$Minute, temp2[,3], col=2, pch=3)
lines(temp2$Minute, temp2[,3], col=2, lty=2)
                LinsVL<-as.data.frame(c(unique(temp$Video_Line_Name), Rs [i],Rs[j],round(tmp.ccc\$rho.c[1],2)))
names(LinsVL)<-c("VideoLine", "Counter1","Counter2","LinsCCC")
Lins<-as.data.frame(rbind(LinsVL,Lins))
}}}}

```


Video Line \(=176\); Lin's CCC \(=0.29\)





Video Line \(=181\); Lin's CCC \(=0.75\)



Video Line = 183 ; Lin's CCC = 0.1







Video Line = 194 ; Lin's CCC = 0.38




10




Video Line \(=210\); Lin's \(C C C=\mathbf{- 0 . 0 7}\)



Video Line \(=210\); Lin's CCC \(=0.08\)





Video Line \(=216\); Lin's CCC = 0.67



Video Line = 223 ; Lin's CCC = 0.38




Video Line \(=223\); Lin's CCC \(=0.12\)








\(14\)




Video Line = 240 ; Lin's CCC = 0.72



Video Line \(=240\); Lin's CCC \(=0.83\)




Video Line \(=253\); Lin's CCC \(=0.86\)


JD_V






Video Line \(=256\); Lin's CCC \(=0.52\)




Video Line \(=259\); Lin's CCC \(=1\)



Video Line \(=259\); Lin's CCC \(=0.98\)







Video Line \(=259\); Lin's CCC \(=1\)






Video Line = 1850 ; Lin's CCC = 0.74





\section*{Lin's CCC results}
- 2/19 station was below the cutoff counts, where Lin's cannot be used on stations with very low counts
- \(5 / 16\) stations passed LINS CCC threshold (0.5) with the first 2 counters.
- \(10 / 16\) stations required a third review where 7 of these stations all 3 reviewers data was used to calculate final density. This is the standard practice on UWTV surveys by the Marine Institute.

Now we will remove the 3rd reviewer from the other stations (3 stations).
```


## stn 210 SOB = 19

## stn 256 JD = 18

## stn 1850 SOB = 19

Recounts1 <- Recounts[-which(Recounts$Video_Line_Name==210 & Recounts$VideoOperatorID==19),]
Recounts1 <- Recounts1[-which(Recounts1$Video_Line_Name==256 & Recounts1$VideoOperatorID==18),]
Recounts1 <- Recounts1[-which(Recounts1$Video_Line_Name==1850 & Recounts1$VideoOperatorID==19),]

```

\section*{Compare percentage difference between the 2019 original (sea) counts vs. the validation (lab) counts}
```


# calculate average count per station for historical review

rec.1 <- Recounts1 %>% group_by(Video_Line_Name, Minute) %>% summarise(ct = mean(BurrowCount))
val.data <- rec.1 %>% group_by(Video_Line_Name) %>% summarise(av.count = sum(ct))
val.data\$method <- "review"

# next extract sea counts (onboard reviewers from 2019)

orig.data <- subset(rec, ! VideoOperatorID %in% c(18,19,20,21))
orig.minute <- with(orig.data, aggregate(BurrowCount, by=list(Minute, Video_Line_Name), FUN=mean))
names(orig.minute) <- c("Minute", "Video_Line_Name", "ct")
orig.data <- with(orig.minute, aggregate(ct, by=list(Video_Line_Name), FUN=sum))
names(orig.data) <- c("Video_Line_Name", "av.count")
orig.data\$method <- "original"
final2019 <- rbind(orig.data, val.data)

# merge both datasets

wide <- merge(orig.data, val.data, by="Video_Line_Name")
wide <- wide[,-c(3,5)]
wide <- rbind(wide, c("Total", colSums(wide) [2:3]))
names(wide) <- c("stn", "orig.count","valid.count")
wide[,2] <- as.numeric(wide[,2])
wide[,3] <- as.numeric(wide[,3])
wide\$perc.change <- with(wide, (valid.count/orig.count)-1)

```

\section*{Barplot of the 2019 original counts and the validation counts by station.}
```

knitr::kable(wide[,1:4])

```
\begin{tabular}{lrrr}
\hline stn & orig.count & valid.count & perc.change \\
\hline 173 & 5.5 & 6.50000 & 0.1818182 \\
176 & 15.0 & 27.00000 & 0.8000000 \\
181 & 20.5 & 29.00000 & 0.4146341 \\
183 & 10.5 & 19.00000 & 0.8095238 \\
194 & 91.5 & 58.00000 & -0.3661202 \\
198 & 6.5 & 7.00000 & 0.0769231 \\
204 & 28.0 & 43.00000 & 0.5357143 \\
209 & 14.0 & 21.00000 & 0.5000000 \\
210 & 6.0 & 20.00000 & 2.3333333 \\
216 & 14.5 & 17.00000 & 0.1724138 \\
223 & 30.5 & 21.66667 & -0.2896175 \\
225 & 8.5 & 14.00000 & 0.6470588 \\
232 & 64.0 & 47.50000 & -0.2578125 \\
240 & 28.0 & 46.50000 & 0.6607143 \\
253 & 11.0 & 30.00000 & 1.7272727 \\
256 & 27.0 & 25.50000 & -0.0555556 \\
258 & 14.0 & 11.00000 & -0.2142857 \\
259 & 11.5 & 15.75000 & 0.3695652 \\
1850 & 19.0 & 32.00000 & 0.6842105 \\
Total & 425.5 & 491.41667 & 0.1549158 \\
\hline
\end{tabular}
```

mean(wide\$perc.change)

## [1] 0.4442353

ggplot(final2019, aes(x=as.factor(Video_Line_Name), y=av.count, fill = method, col=method))+
geom_bar(width=0.8, stat="identity",position=position_dodge())+
xlab("Station ID") + ylab("count")+
theme_bw()

```


Figure 2: Bar plot showing 2019 original (onboard) counts and the validation (lab) counts

\title{
Violin plot of the 2019 original counts and the validation counts.
}
```

\#final2016 <- cbind(final2019, wide[-nrow(wide),])
ggplot(final2019,aes(x=as.factor(method),y=av.count))+
geom_violin(aes(group=method,colour=method,fill=method), alpha=0.5,
kernel="rectangular")+ \# passes to stat_density, makes violin rectangular
geom_boxplot(aes(group=method), width=.2)+
stat_summary(fun.y=mean, geom="line", colour="blue", aes(group=1)) +
xlab("method")+ \# label one axis
ylab("av.count")+ \# label the other
theme_bw()+ \# make white background on plot
theme(legend.position = "none") \# suppress legend

```


Figure 3: Violin and box plot of counts distributions of 2019 original counts and validation counts. The blue line indicates the mean density over time. The horizontal black line represents the median, white box is the inter quartile range, the black vertical line is the range and the black dots are outliers.

\section*{UWTV camera calibration test. Marine Institute}

\section*{Introduction}

The Marine Institute (MI) has been carrying out UWTV surveys since 2002 to estimate Nephrops norvegicus abundances in several Functional Units (FU). These surveys are key to the stock assessment of Nephrops. In order to use the latest technology available, in 2019 the MI replaced the standard definition camera (SDc) used in the last years with a new high definition camera (HDc). This document presents the results of a calibration test conducted by the MI to compare burrow counts from both cameras.

\section*{Material \& Methods}

14 stations were recorded with both cameras at the same time during the Porcupine bank Nephrops grounds (FU16) 2019 UWTV survey (Aristegui et al., 2019). Both cameras were mounted in the same sledge used in previous UWTV surveys: the SDc was set up as in previous surveys at an angle of \(40^{\circ}\) to the bottom, while the HDc was set up at an angle of \(75^{\circ}\) (Table 1). In each station 10 minutes of good quality footage were recorded by each camera, assuming that both cameras recorded exactly the same track of seafloor.

Table 1. UWTV camera calibration test. Features of the two camera systems.
\begin{tabular}{|c|c|c|}
\hline & Standard Definition camera & High Definition camera \\
\hline \begin{tabular}{c} 
Camera angle \\
to the bottom
\end{tabular} & \(40^{\circ}\) & \(75^{\circ}\) \\
\hline \begin{tabular}{c} 
Field of View \\
(FOV)
\end{tabular} & 0.75 m & 1.01 m \\
\hline Footage format & DVD & Digitalized stills (12 frames per second) \\
\hline \begin{tabular}{c} 
Counting \\
method
\end{tabular} & Hand writing time stamped & \begin{tabular}{c} 
Image annotation R Shiny app \\
(Aristegui, 2019)
\end{tabular} \\
\hline
\end{tabular}

The HDc footage was counted at sea by five trained scientists using an inhouse developed image annotation R Shiny app (Aristegui, 2019). The SDc footage was counted back at the MI office by four of the five scientists who counted the HDc, using the same method as in previous FU16 UWTV surveys (hand writing the time stamp of each burrow). The 14 stations from each camera were assigned randomly and equally to the scientist team. Each station was counted independently by two scientists.

Both SDc and HDc count data were analysed in the same way independently one from the other. The counts were screened to check for any unusual discrepancies using Lin's Concordance Correlation Coefficient (CCC) with a threshold of 0.6 (Lin, 1989). Those stations that did not pass the threshold were counted by a third scientist.

Count data that passed the threshold were averaged in order to get a mean burrow count per minute for each of the 28 stations. As the cameras differ in their field of views (FOV) (Table1), the counts were standardized dividing them by their corresponding FOV. Finally, a paired t-test was used to compare both datasets.

\section*{Results}

The standardized counts for both methods were in a similar range of burrows per minute divided by FOV: from 0.4 to 7.9 for SDc, and from 0.4 to 6.7 for HDc (Figure 1). The conducted test suggests that, in average, there is not significance difference between the two methods ( \(p\)-value \(=0.06563>0.05\) ).


Figure 1. UWTV camera calibration test. Standardized counts of each station and boxplots. Standard Definition camera counts (left) and High Definition camera counts (right). Same stations are linked with a dotted line.

\section*{Conclusion}

The independent results of each camera system are very similar, and the new HD camera performed appropriately, in line with the SD camera used in UWTV surveys by the MI until 2018.

On top of the obvious better quality of the footage, the HDc system also allows a smoother workflow onboard, as all the footage is now digitalized. Together with the image annotation app, the HDc system makes the survey process paperless and less prone to errors, as there is no need of inputting manually the count data into the databases any more.

\section*{References}

Aristegui, M. (2019) Image annotation R Shiny app. Marine Institute. http://doi.org/c8jt
Aristegui, M., O’Brien, S., Tully, D., Galligan, S., McCorriston, P., Bentley, K. and Lordan, C. (2019). Porcupine Bank Nephrops Grounds (FU16) 2019 UWTV Survey Report and catch scenarios for 2020. Marine Institute UWTV Survey report.

Lin, L. I-K. (1989). A Concordance Correlation Coefficient to Evaluate Reproducibility. Biometrics, 45(1), 255-268. doi:10.2307/2532051

\section*{SIAMISS Estimates of Anglerfish Biomass in Subareas IV and VI for 2019}

\author{
Elisa Barreto, Liz Clarke, Gerald McAllister, James Dooley, Ruadhan Gillespie-Mules, Eoghan Kelly, Hans Gerritsen \\ Marine Scotland Science \& Marine Institute, Ireland
}
- We report the anglerfish results for the Scottish Irish Anglerfish Megrim Industry Science Survey (SIAMISS) in 2019. ICES have used this survey as the basis for advice under the data-limited approach for category 3 stocks since 2012.
- The estimated biomass for 2019 in ICES Subareas IV \& VI is 59 kt ( \(95 \% \mathrm{CI}: 46\) - 71kt), a decrease of \(25 \%\) from 78 kt in 2018, and \(34 \%\) to 2017 levels.
- The average biomass over the last two years \((2018-2019)\) is 68 kt , compared to an average biomass of 78 kt in the previous 3 years ( \(2015-2017\) ). This is a decrease in average biomass (last two years versus previous three years) of \(13 \%\) ( \(95 \%\) CI: \(1 \%-25 \%\) ).

\section*{Introduction}

Anglerfish are a group of fish of the order Lophiiformes which include several families, amongst which the Lophiidae comprise some of the important commercially fished species of the genus Lophius (Fernandes et al. 2007). Lophius piscatorius and Lophius budegassa (black-bellied anglerfish) are more commonly known collectively as "monkfish". In 2017, 12.944 tonnes of monkfish were landed into Scotland by UK vessels, with a value of \(£ 36.80\) million (Anon, 2018).

This document focuses on the anglerfish that occur on the Northern Shelf. They are considered and assessed as one stock covering ICES Division IIIa (Skagerrak \& Kattegat), ICES Subarea IV (North Sea), and ICES Subarea VI (West of Scotland and Rockall). In 2005, Fisheries Research Services (FRS, now Marine Scotland Science - MSS) started a new survey to estimate the abundance and distribution of anglerfish on the Northern Shelf. Initially, and again in recent years, the survey has included contribution from research vessel of the Marine Institute in Ireland and is called the Scottish Irish Anglerfish Megrim Industry Science Survey (SIAMISS). This survey covers much of the area of the known distribution of northern shelf anglerfish (ICES Divisions IVa, VIa and VIb at Rockall), with the exception of the central and southern parts of Subarea IV and the Skagerrak and Kattegat (Division IIIa).

Because it covers such a large area, the current design incorporates multiples vessels to survey the whole area.

\section*{Material and Methods}

The survey area covers the northern shelf of the British Isles, north of latitude \(56^{\circ}\) to a northerly limit of \(62^{\circ} 30^{\prime}\) north. This area was limited to zones where the depth was less than 1000 m (trawling only occurred in waters up to a maximum of 1000 m ).

The survey area was stratified based on knowledge from fishermen (Fig. 1). The sampling effort within each stratum was allocated roughly according to its expected biomass, and the sample locations were chosen at random from grids of points within strips of equal area. This is to ensure equal probability of selection and even coverage within a stratum (ICES 2013). The aim was for a minimum of 3 stations within a stratum.

The 2019 anglerfish multi-vessel survey took place from \(28^{\text {th }}\) April to \(7^{\text {th }}\) May 2019 and involved two vessels: Scottish commercial fishing vessel MFV Genesis - surveying Division IVa, Division VIa North of \(58^{\circ} \mathrm{N}\), and Rockall and the Irish Marine Institute research vessel FRV Celtic Explorer, surveying Division VIa South of \(58^{\circ} \mathrm{N}\). One haul with the duration of 60 minutes was made at each sampling station ( \(\mathrm{n}=128\) ). Each vessel on the survey employed exactly the same gear, the specification of which was drawn up in partnership with industry. Until 2017, every anglerfish caught was measured for length, sex, maturity, total and gutted weight, and the otolith was removed, however due to increasing numbers caught, from 2018, a length-stratified approach to sampling for sex, maturity, weight and age was applied

All Scottish cruises used a Scanmar system to monitor trawl parameters such as wing spread, door spread and headline height. A self-recording tilt meter rigged at ground gear centre monitored touch-down/lift-off at block up/knockout and to check if the ground gear lifted off during the haul. For the FRV Celtic Explorer, door spread, wing spread, headline height and bottom contact were monitored using Scanmar and Marport trawl sensors (distance sensors in the doors and wing-ends, headline sensor and a trawl-eye sensor positioned on the topsheet directly over the footrope) (Gerritsen et al., 2018).

For the Scottish data, the swept areas of the wings and doors were then calculated as the sum of the individual wing/door spread measurements multiplied by the distance travelled between successive measurements (Fernandes et al., 2007). For the Irish data, the median values of the door spread, wing spread and headline height were recorded at the end of the
tow. These measurement as well as bottom depth and GPS position are recorded in a SQL database at intervals of approximately 1 per second (Gerritsen et al., 2018).

The estimation of abundance for anglerfish takes into account the following factors:
1) herding of anglerfish by the trawl doors and sweeps;
2) escapes of fish under the trawl footrope;
3) anglerfish abundance and biomass in the southern part of Division VIa not covered in 2005, 2008 and 2010;
4) variability due to: a) sampling, b) herding (based on experimental data), c) footrope escapes (based on experimental data).

Herding corrections were based on a model derived from observations of anglerfish behaviour using video cameras mounted on the sweeps: full details are described in Reid et al (2007a). The number of fish escaping under the footrope has been estimated from experimental data using catching bags under the footrope (Reid et al 2007b). The number and size of anglerfish passing under and into these bags were measured. A size based model of footrope selectivity was then developed. This model was then applied to the length data from each survey to correct for those fish that were likely to escape under the net (ICES 2009). Prior to 2011, Scotland only surveyed the Northern part of ICES Division VIa, with this area being covered by Ireland in 2006, 2007 and 2009. Estimates of the proportion of anglerfish in the southern part of ICES Division VIa for 2005, 2008 and 2010 were derived from 2006, 2007 and 2009 when Ireland contributed to the survey and covered this area completely, using averages of the proportions of the abundance in this area relative to the whole Northern shelf. The averages of these proportions \((8.7 \%\) for abundance and \(5.1 \%\) for biomass) were used to estimates of the surveys in 2005, 2008 and 2010 when the Irish did not participate. Since 2011, Scotland has surveyed the whole of ICES Division VIa.

The estimates currently do not take account of the following:
- areas in the central and southern North Sea (eastern part of ICES Division IVa and all of IVb and IVc);
- areas inaccessible to the trawl in Division VIa.

Further details of the survey methods are given in Fernandes et al. (2007). Also for a full description of the estimation of abundance and catchability components see Annex 2 - Stock Annex in ICES (2013).

\section*{Results}

Figures 2 and 3 show the survey haul locations and mean numbers and weight per \(\mathrm{km}^{2}\) caught at these locations. Larger numbers of anglerfish were caught along the shelf-edge below \(58^{\circ} \mathrm{N}\), with large weights of fish being caught at the same locations and also at Rockall, indicating that the fish at Rockall are larger than those caught on the shelf-edge. This is reflected in the estimated population total numbers-at-length and total weight-at-length by Division (Figure 4), which show small fish ( \(\sim 20 \mathrm{~cm}\) ) in Divisions IVa and VIa, and a much higher proportion of large fish at Rockall. Comparison of numbers-at-length and weight-atlength over time show a larger proportion of larger fish in all areas compared to recent years (Figures 5-8).

Tables 1 - 3 give point and variance estimates for total numbers and biomass for 2019 and previous years, by Division and for Subareas IV \& VI combined. Biomass in Divisions IVa and VIa had a slight decrease, whilst that in VIb had a substantial drop compared to last year. (Table 2, Figure 9). The total biomass estimate for the survey area for 2019 is 58.575 kt , an overall decrease of \(25 \%\) compared to 2018 . Although this value of biomass is the lowest since 2014, it is in line with the values from previous years (2005-2013).

The average biomass over the last two years (2018-2019) was 68.118 Kt , compared to an average biomass of 77.919 kt in the previous 3 years (2015-2017). This leads to a decrease in average biomass (last two years versus previous three years) of \(13 \%\) ( \(95 \%\) CI: \(1 \%-25 \%\) ) (Table 4).

The percentage of the biomass in area IV compared to the biomass for areas IV and VI combined has oscillated around 50\% for the years 2005-2018 (Figure 11). This is in contrast to the division of the TAC across ICES subareas IV and VI, which has been in the ratio 64:36 (IV:VI) since 2011.

\section*{References}

Anon. (2018). Scottish Sea Fisheries Statistics 2017. Scottish Government, Edinburgh, 104pp.

Fernandes, P.G., Armstrong, F., Burns, F., Copland, P., Davis, C., Graham, N., Harlay, X., O’Cuaig, M., Penny, I., Pout, A.C. and Clarke, E.D. (2007). Progress in estimating the absolute abundance of anglerfish on the European northern shelf from a trawl survey. ICES CM 2007/K: 12.16 pp.

Gerritsen, H.D, Kelly, E., Stokes, D., O'Hea, B., \& Ni Chonchuir, G. (2018). Cruise report: Irish Anglerfish \& Megrim Survey 2017. FEAS Survey Series: IAMS 2017. Marine Institute.
ICES. 2013. Report of the Working Group for Celtic Seas Ecoregion (WGCSE), 8-17 May 2013, Copenhagen, Denmark. ICES CM 2013/ACOM: 12. 1986 pp.
Reid, D. G., Allen, V. J., Bova, D. J., Jones, E. G., Kynoch, R. J., Peach, K. J., Fernandes, P. G., and Turrell, W. R. (2007). Anglerfish catchability for swept-area abundance estimates in a new survey trawl. ICES Journal of Marine Science 64:1503-1511.
Reid, D.G., Kynoch, R.J., Penny, I. and Peach, K. (2007). Estimation of catch efficiency in a new angler fish survey trawl. ICES CM 2007/Q: 22.20 pp.

\section*{Acknowledgements}

Thanks to all the skippers and crew of the vessels used this year and on previous surveys, and to the MSS, SFF and MI staff who have participated in the surveys or analysed the data over the years.

Table 1: Abundance of anglerfish from the 2019 survey by region. CI= Confidence Interval; RSE \(=\) relative standard error.
\begin{tabular}{|lr|rrrr|rrrr|}
\hline Sector & Area & \begin{tabular}{c} 
Numbers \\
(millions)
\end{tabular} & \begin{tabular}{c} 
CI \\
lower
\end{tabular} & \begin{tabular}{c} 
CI \\
upper
\end{tabular} & RSE & \multicolumn{1}{c|}{\begin{tabular}{c} 
Biomass \\
(kt)
\end{tabular}} & \begin{tabular}{c} 
CI \\
lower
\end{tabular} & \begin{tabular}{c} 
CI \\
upper
\end{tabular} & RSE \\
\hline \begin{tabular}{l} 
Division IVa \\
(partial)
\end{tabular} & 159828 & 14.606 & 11.758 & 17.454 & 9.948 & 23.719 & 17.206 & 30.232 & 14.010 \\
\hline Division VIa & 116307 & 21.032 & 5.885 & 36.179 & 36.746 & 18.864 & 9.004 & 28.723 & 26.667 \\
\hline Division VIb & 39567 & 3.592 & 2.785 & 4.399 & 11.460 & 15.992 & 12.277 & 19.707 & 11.852 \\
\hline Subarea VI & 155874 & 24.624 & 9.456 & 39.793 & 31.430 & 34.856 & 24.320 & 45.392 & 15.423 \\
\hline IV \& VI & 315703 & 39.231 & 23.797 & 54.664 & 20.073 & 58.575 & 46.189 & 70.962 & 10.789 \\
\hline
\end{tabular}

Table 2: Survey estimates of anglerfish abundance for 2005-2019, by region.
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline \multirow{2}{*}{Year} & \multirow{2}{*}{Month} & \multicolumn{5}{|c|}{Numbers (millions)} & \multicolumn{5}{|c|}{Biomass (kt)} \\
\hline & & IVa & VIa & VIb & VI & Total & IVa & VIa & VIb & VI & Total \\
\hline 2005 & November & 11.168 & 10.866 & 1.800 & 12.666 & 23.834 & 18.642 & 14.096 & 5.879 & 19.975 & 38.617 \\
\hline 2006 & November & 12.844 & 10.459 & 3.174 & 13.633 & 26.477 & 21.921 & 12.175 & 6.889 & 19.064 & 40.985 \\
\hline 2007 & November & 15.304 & 7.956 & 4.000 & 11.956 & 27.26 & 28.534 & 11.072 & 10.786 & 21.858 & 50.392 \\
\hline 2008 & April & 12.613 & 7.718 & 3.952 & 11.67 & 24.283 & 29.721 & 14.383 & 9.442 & 23.825 & 53.546 \\
\hline 2009 & April & 8.279 & 5.144 & 3.688 & 8.832 & 17.111 & 17.058 & 8.150 & 12.852 & 21.002 & 38.060 \\
\hline 2010 & April & 7.366 & 5.161 & 3.131 & 8.292 & 15.658 & 21.944 & 11.59 & 8.745 & 20.335 & 42.279 \\
\hline 2011 & April & 5.150 & 6.057 & 3.669 & 9.726 & 14.876 & 14.949 & 9.330 & 8.974 & 18.304 & 33.253 \\
\hline 2012 & Abril & 5.432 & 4.961 & 5.135 & 10.096 & 15.528 & 15.106 & 9.213 & 12.005 & 21.218 & 36.325 \\
\hline 2013 & October & 8.470 & 8.461 & 4.885 & 13.346 & 21.816 & 14.369 & 10.801 & 13.626 & 24.427 & 38.796 \\
\hline 2014 & April & 17.553 & 16.096 & 6.488 & 22.584 & 40.136 & 21.284 & 16.633 & 14.967 & 31.60 & 52.884 \\
\hline 2015 & April & 18.266 & 28.604 & 5.496 & 34.100 & 52.366 & 29.653 & 24.047 & 14.215 & 38.262 & 67.915 \\
\hline 2016 & April & 21.648 & 14.383 & 4.538 & 18.922 & 40.569 & 43.956 & 18.273 & 15.717 & 33.99 & 77.946 \\
\hline 2017 & April & 23.691 & 16.332 & 4.360 & 20.683 & 44.374 & 46.995 & 29.297 & 11.604 & 40.901 & 87.896 \\
\hline 2018 & April & 11.819 & 13.528 & 6.240 & 19.768 & 31.586 & 29.353 & 22.350 & 25.958 & 48.308 & 77.661 \\
\hline
\end{tabular}

Table 3: Total biomass of anglerfish (kt) in Subareas IV \& VI estimated from the surveys for 2005-2019. RSE \(=\) relative standard error; \(\mathrm{CI}=\) Confidence Interval.
\begin{tabular}{|c|crrrrr|}
\hline Year & \begin{tabular}{c} 
Number \\
hauls
\end{tabular} & \begin{tabular}{c} 
Number \\
measured
\end{tabular} & Biomass & RSE & \begin{tabular}{r} 
CI \\
lower
\end{tabular} & \begin{tabular}{r} 
CI \\
upper
\end{tabular} \\
\hline 2005 & & & 38.617 & 20.000 & 23.479 & 53.755 \\
2006 & & & 40.985 & 8.100 & 34.478 & 47.492 \\
2007 & 156 & 1569 & 50.392 & 6.583 & 43.676 & 57.108 \\
2008 & 167 & 2219 & 53.546 & 7.338 & 42.421 & 64.671 \\
2009 & 206 & 1643 & 38.060 & 6.555 & 32.987 & 43.133 \\
2010 & 168 & 1280 & 42.279 & 8.583 & 30.429 & 54.129 \\
2011 & 153 & 1037 & 33.254 & 7.466 & 24.846 & 41.662 \\
2012 & 169 & 1461 & 36.325 & 9.551 & 29.704 & 42.946 \\
2013 & 93 & 984 & 38.395 & 11.526 & 31.020 & 45.770 \\
2014 & 106 & 1568 & 52.884 & 9.275 & 42.769 & 62.999 \\
2015 & 117 & 2198 & 67.915 & 6.861 & 58.782 & 77.047 \\
2016 & 108 & 2025 & 77.946 & 7.275 & 66.831 & 89.060 \\
2017 & 153 & 3265 & 87.896 & 7.937 & 74.222 & 101.569 \\
2018 & 142 & 2714 & 77.661 & 7.491 & 66.258 & 89.064 \\
2019 & 128 & 1860 & 58.575 & 10.789 & 46.189 & 70.962 \\
\hline
\end{tabular}

Table 4: Average biomass 2018-2019 (last two years) compared to average biomass 20152017 (previous 3 years).
\begin{tabular}{llrl}
\hline Sector & Period & Average Biomass Estimate (kt) & \% change \\
\hline \multirow{2}{*}{ IV \& VI } & \(2015-2017\) & 77.919 & \multirow{2}{*}{\(13 \%(95 \% \mathrm{CI}: 1 \%-25 \%)\)} \\
& \(2018-2019\) & 68.118 & \\
\hline
\end{tabular}

a)

b)

Figure 1: Map of the northern continental shelf around Scotland showing the sectors (a) and stratums (b) with haul locations, covered by the three survey cruises 2019.


Figure 2: Map of the northern continental shelf around Scotland showing the density of anglerfish (in numbers) caught in SIAMISS 2019. Each blue circle is centred on the sample location and circle size is proportional to the number density caught per haul in \(\mathrm{n} / \mathrm{km}^{2}\) according to the legend (top left). The green lines represent the coastline, the grey lines the depth contours, and the black lines the stratum boundaries.


Figure 3: Map of the northern continental shelf around Scotland showing the density of anglerfish (in terms of biomass) during SIAMISS 2019. Each red circle is centred on the sample location and circle size is proportional to the density caught per haul in \(\mathrm{kg} / \mathrm{km}^{2}\) according to the legend (top left). The green lines represent the coastline, the grey lines the depth contours and the black lines the stratum boundaries.


Figure 4: SIAMISS estimates of anglerfish numbers-at-length (left) and weight-at-length (right) by region for 2019.


Figure 5: SIAMISS estimates of total numbers-at-length (left) and total weight-at-length (right) by year (2007-2019) for all areas combined.


Figure 6: SIAMISS estimates of total numbers-at-length (left) and total weight-at-length (right) by year (2007-2019) for Division IVa.


Figure 7: SIAMISS estimates of total numbers-at-length (left) and total weight-at-length (right) by year (2007-2019) for Division VIa. (Note the different scale on numbers-at-length for \(2014 \& 2015\).)


Figure 8: SIAMISS estimates of total numbers-at-length (left) and total weight-at-length (right) by year (2007-2019) for Division VIb.


Figure 9: SIAMISS estimates of total numbers (left) and biomass (right) for anglerfish by region and year (2005-2019).


Figure 10: SIAMISS anglerfish biomass estimates for 2005-2019. The horizontal grey lines represent the average biomass for 2015-2017 and the average biomass for 2018-2019.


Figure 11: Biomass from ICES subarea IV as a percentage of biomass from ICES subareas IV and VI combined for 2005-2019. The dashed line shows the mean percentage across the years ( \(48.32 \%\) ), the dotted lines show the percentage used for setting TACs in the two areas.

\section*{Annex 4: Technical Minutes}

The following pages consist of the Review of the relevant stocks conducted at the University of Maine Orono, Maine, USA from May 28-June 7, 2019/


\title{
Technical Minutes of the University of Maine Review Group for the Advice Drafting Group for the Celtic Seas
}

May 28- June 7, 2019
University of Maine Orono, Maine, USA

\section*{Reviewers:}

Mackenzie Mazur (Chair), Dr. Bai Li (Co-Chair), Luoliang Xu (Co-Chair), Ming Sun (CoChair), Cameron Hodgdon (Co-Chair), Jamie Behan, Bowen Chang, Hsiao-Yun Chang, Ning Chen, Marina Cucuzza, Libin Dai, Yanan Li, Yunzhou Li, Robyn Linner, Dr. Qiuyun Ma, Shu Su, Harriett Train, Jaelee Vanidestine, Jiaqi Wang, Nathan Willse, Lei Xing, and Qilei Zhao

Faculty Advisor:
Dr. Yong Chen (Professor, School of Marine Sciences, University of Maine)

\section*{ICES Secretariat:}

Liese Carleton

\section*{Review Process}


The University of Maine Review Group (UMaine RG) met on May 28th, 2019 to examine the review materials, discuss the review process, and assign individuals to a subgroup of 4-5 reviewers focusing on a particular stock(s). The relevant materials were distributed to each RG subgroup when they became available on May 31st on the ICES SharePoint website. Reviews were carried out after the working group (WG) completed the final report for WGCSE and HAWG stocks. In general, the ICES guidelines for review groups (RG) were followed. The RG focused on the consistency between the WG report and the stock annex, i.e., checking whether the assessment, calculation of biological reference points (BRPs), and forecast were carried out in accordance to stock annex, Terms of reference (ToRs), and RG guidelines. Furthermore, the RG examined the data quantity and quality, assessment method, technical measures, uncertainty, and BRPs for each reviewed stock to ensure that management measures are based upon the best scientific information available. The RG finalized their reports on June 7th to determine the status of each group's report as well as their final decision of accepting or rejecting the assessment and discuss any remaining issues. Table 1 lists the stocks reviewed by the UMaine RG along with the suggestion (accept, accept with caveats, or reject).

Table 1. List of stocks reviewed by the University of Maine RG.
\begin{tabular}{|l|l|l|l|l|}
\hline \begin{tabular}{l} 
Stock code/draft \\
advice link
\end{tabular} & Assessment method & \begin{tabular}{l} 
Data \\
category
\end{tabular} & \begin{tabular}{l} 
Expert \\
group
\end{tabular} & RG suggestion \\
\hline bss.27.4bc7ad-h & SS3 & 1 & WGCSE & Accept with caveats \\
\hline cod.27.6a & TSA & 1 & WGCSE & Accept with caveats \\
\hline cod.27.7a & TSA & 3 & WGCSE & Accept with caveats \\
\hline cod.27.7e-k & XSA & 3 & WGCSE & Accept with caveats \\
\hline had.27.7b-k & ASAP & 1 & WGCSE & Accept with caveats \\
\hline lez.27.4a6a & \begin{tabular}{l} 
Bayesian state-space surplus \\
production model
\end{tabular} & 1 & WGCSE & Accept with caveats \\
\hline ple.27.7a & SAM & 1 & WGCSE & Accept with caveats \\
\hline ple.27.7e & XSA & 3 & WGCSE & Accept \\
\hline ple.27.7fg & SPiCT & 3 & WGCSE & Accept with caveats \\
\hline ple.27.7h-k & XSA & 3 & WGCSE & Accept with caveats \\
\hline pol.27.67 & DCAC & 4 & WGCSE & Accept with caveats \\
\hline sol.27.7a & XSA & 3 & HAWG & Accept \\
\hline sol.27.7e & XSA & 1 & WGCSE & Accept with caveats \\
\hline sol.27.7fg & XSA & 1 & WGCSE & Accept \\
\hline sol.27.7h-k & XSA & 1 & WGCSE & Accept with caveats \\
\hline whg.27.7a & ASAP & 1 & WGCSE & Accept with caveats \\
\hline whg.27.7b-ce-k & XSA & 1 & WGCSE & Accept with caveats \\
\hline pok.27.7-10 & DCAC & 1 & WGCSE & Accept \\
\hline her.27.6a7bc & SAM & Hccept with caveats \\
\hline her.27.irls & ASAP & 1 & Hccept with caveats \\
\hline her.27.nirs & FLSAM & 1 & HAWG & Accept with caveats \\
\hline spr.27.67a-cf-k & N/A & HAWG & Accept with caveats \\
\hline spr.27.7de & Biomass trend & 1 & Accept with caveats \\
\hline
\end{tabular}

\section*{Table of Contents}
Review process ..... 2
Table 1. List of stocks reviewed by the University of Maine RG ..... 3
General comments ..... 5
Stock specific issues ..... 7
Individual review reports:
bss.27.4bc7ad-h. ..... 10
cod.27.6a ..... 14
cod.27.7a ..... 17
cod.27.7e-k ..... 19
had.27.7b-k ..... 24
lez.27.4a6a ..... 26
ple.27.7a ..... 31
ple.27.7e ..... 35
ple.27.7fg ..... 38
ple. \(27.7 \mathrm{~h}-\mathrm{k}\) ..... 40
pol.27.67 ..... 45
sol.27.7a. ..... 48
sol.27.7e ..... 52
sol.27.7fg ..... 55
sol.27.7h-k ..... 58
whg. 27.7 a . ..... 61
whg.27.7b-ce-k ..... 67
pok.27.7-10 ..... 71
her.27.6a7bc ..... 74
her.27.irls ..... 78
her.27.nirs ..... 82
spr.27.67a-cf-k ..... 87
spr.27.7de ..... 89
References ..... 94

\section*{General Comments}

The RG identified some common issues regarding the reports, data, models, and BRPs among the stocks, and they are summarized as follows:

The RG recommends clearly justifying why some stocks are combined and some are not. Often, the descriptions of stock area need further justification.

The RG recommends trying multiple scenarios regarding landing data: one with ICES adjusted landings and one with official landings, since these landings values often differ.

When making assumptions about discards, it would be beneficial to provide detailed justification for the assumptions, as they can have a large effect on the assessment results and modeling residuals. The RG commends the WG for utilizing discard data, but sometimes the assumptions about discards in years where discard data are not available are not clear. Additionally, the RG recommends running structured sensitivity analyses with discard data and with different discard rate assumptions.

Ecosystem and growth rate changes were a common thread to many of the reports. Thus, the RG was disappointed that few assessments considered a structured sensitivity analysis for natural mortality (Butterworth and Rademeyer 2008). Natural mortality is a major source of uncertainty in all fishery stock assessment and with the ecosystem changes present in these regions, it is especially important to explore how growth and predation-mediated changes in M affect determination of stock status and projection. In instances where \(M\) is expected to change, the RG suggests exploration of M-RAMP or time block natural mortality (Brooks et al. 2016, Legault and Palmer 2016). More thorough knowledge will inform the WG of changes in stock productivity. This may be particularly crucial to stocks that have seen distributional or growth shift in recent years.

The RG recommends presenting parameters and their justifications in a table, which would make the reports easier to read. Providing figures and a table to present results from sensitivity analyses would also benefit the reports.

Quantifying uncertainty of the estimates of biomass and recruitment would be beneficial to the assessments. Additionally, the XSA assessments would benefit from a structured framework for incorporating uncertainties. A paper by Scott et al. (2015) provides such a framework: 1) generating the candidate assumptions, 2) estimating parameters of the stock assessment model, and 3 ) averaging across the model results.

It would be helpful for readers if the relationships between BRPs were summarized in the assessment reports. The RG would like to know how Fpa relates to Fmsy. Also, the harvest control rules (HCRs) based on these reference points should be described.

The RG appreciates that retrospective analysis was conducted for most stocks as consistent with the ToRs. The RG agrees with the ICES advice and concur it should be routinely checked in the assessment process. However, it is unclear how many years were peeled in the retrospective analysis. The RG also recommends that the number of retrospective peels should be justified and should reflect the average lifespan of recruited individuals in the stock area. For example, there should be more than a 5 year peel for retrospective analyses for cod stocks. The RG also suggests that a guideline be developed to correct retrospective errors in the determination of stock status and projection.

As many of the stocks are driven by high recruitment events that are usually not highly related to the spawning stock biomass (SSB), the RG recommends exploring relationships between recruitment and environmental variables. This may fill in large gaps in information about recruitment. Also, for stocks with highly variable recruitment due to environmental conditions, recruitment scenarios in the projections should not only include the geometric mean of recruitment values but also reflect high and low recruitment events that could be related to environmental scenarios.

\section*{Stock Specific Issues}

The RG suggested the following stocks to be accepted if certain suggestions are considered (accept with caveats). These stocks and suggestions are:
bss.27.4bc7ad-h
a) Update the assessment with more recent data (2016-2018)
b) Calculate Mohn's Rho
cod.27.6a
a) Provide more discussion on Cook (2019)'s methods, the uncertainties associated with TSA, and additional assessment methods
cod.27.7a
a) Provide more information about retrospective bias
b) Include complete methodology of the new method in the assessment report or update the stock annex
c) Provide more discussion on model comparisons
cod.27.7e-k
a) Address issues regarding the use of survey data
b) Address issues regarding the uncertainty in catch and discard data
c) Account for the high uncertainty in estimates of key parameters by using a stochastic projection
had.27.7b-k
a) Compare the retrospective bias between XSA and ASAP models
b) Research the relationship between environmental variables and recruitment
c) Reconsider recruitment configurations in the projections (i.e. scenarios with high and low recruitment values)
lez.27.4a6a
a) Conduct exploratory runs to capture the uncertainty of total catch by modeling catch with variance corresponding to a constant coefficient of variation
b) Use a generalized additive model (GAM) to derive model-based abundance indices and assign different indices with different weights in the stock assessment
c) Conduct the assessment with a more generalized approach by incorporating a shape parameter to control the level of biomass as a proportion of unfished biomass at which surplus production is maximized
ple.27.7a
a) Update the stock annex so that the reader can better understand the model
b) Evaluate impacts of stochastic effects of natural mortality on the model performance
ple.27.7fg
a) Provide more discussion on the environmental vulnerability of the stock
b) Provide a more detailed introduction for the input data and interpretations of assessment diagnostics
ple. \(27.7 \mathrm{~h}-\mathrm{k}\)
a) Discuss large retrospective bias
b) Conduct exploratory assessment runs to investigate uncertainties, especially in discards, natural mortality, and the stock-recruitment relationship
c) Use the data collected from 7 h to make the outputs of the stock assessment representative of the entire management area
d) Provide a detailed explanation of the stock status and management plan
pol.27.67
a) Collect data on the recreational fishery
b) Explore other assessment models
sol.27.7a
a) Try and cross validate different parameter settings to ensure that the best prediction of the model is being used
b) Conduct a simulation study to test whether the low sample sizes would accurately reflect the size distribution and age structure of sole
sol.27.7fg
a) Conduct a LPUE standardization analysis to improve the reliability of the fishery dependent data
b) Provide more discussion on the consequences of ignoring discard or observation errors associated with the catch data
c) Explore alternative models (e.g., ASAP, SAM)
d) Justify the reference points and clearly describe the HCR
sol.27.7h-k
a) Reevaluate all information, tables, and figures presented in the report and annex to ensure that the information being presented is true for sol. \(27.7 \mathrm{~h}-\mathrm{k}\) and not information accidentally copied from a different stock
b) Consider more conservative BRPs
c) Provide more discussion on the expected effects of the new management plan
whg. 27.7 a
a) Conduct exploratory runs to investigate the sensitivities of results to changes in fishery input data, natural mortality values, and model parameter steepness (h)
b) Provide information on maximum fishing effort that can be expected under management based on Fmsy
c) Remove more years of data in the retrospective analysis
pok.27.7-10
a) Explore a stock assessment model
b) Identify and evaluate data gaps and quality
her.27.6a7bc
a) Provide more information about the existing sensitivity analysis and the difference in settings with the previous assessment
b) Conduct an alternative sensitivity analysis for the unavailable biology data and assumptions for catchability
her.27.irls
a) Address issues regarding the use of survey data
b) Conduct stochastic projections to account for extremely high uncertainty for key parameters
her.27.nirs
a) Research uncertainty related to spatial and temporal variability, stock mixing, natural mortality, and catchability assumptions
b) Describe the retrospective analysis
c) Revise figures and tables
spr.27.67a-cf-k
a) Add context for the 20133500 t limit decision
bss.27.4bc7ad-h: Sea bass (Micromesistius poutasso) in divisions 4.b-c, 7.a, and 7.d-h (central and southern North Sea, Irish Sea, English Channel, Bristol Channel, and Celtic Sea)

\section*{1. Assessment Type: Update}
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- Short term projection - Carried out by Stock Synthesis 3 (SS3) from 2017 to 2019.
- No medium or long term was carried out.
4. Assessment Method: Stock Synthesis 3 (SS3) was used.
5. Consistency:
- There were small changes to the Solent bass survey in recent years to ensure consistency with values from previous years.
- The Channel Ground Fish Survey (CGFS) switched vessels in 2015, and an index was estimated considering this change in catchability. However, since the index requires review, it was not used.
- An extraction from French logbook data in 2019 indicated that discards are underestimated from on-board sampling.
- The model is sensitive to recent data, because the recent data have higher influence given the higher sampling levels.
- The most recent fishing mortality (F) estimates are the least precise, and the F estimate for 2014 could be revised downwards in future assessments.

\section*{6. Stock Status:}
- SSB is estimated to be below \(\mathrm{B}_{\mathrm{pa}}\) but above \(\mathrm{B}_{\mathrm{lim}}\) in 2015.
- The combined commercial and recreational fishery F is well above the FMSY proxy.
- Total landings in 2018 was 912 t.
- Catches decreased from 949 t in 2017.
- SSB increased from the mid 1990's, peaked in 2009 and has declined rapidly in recent years.
- Recruit is at a low-level, fluctuating without trend since 2008.

\section*{7. Management and Biological Reference Points:}
- Total removals in 2019 should be no more than 1789 t .
- \(\mathrm{F}_{\text {MSY }}\) proxy and \(\mathrm{B}_{\text {lim }}\) reference points defined in 2016 have not been altered.
- For the recreational fisheries, minimum sized increased from 36 cm to 42 cm in 2015, and bag limits and closed seasons were imposed.
- No TAC has been defined.
- MSY advice currently uses \(\mathrm{F} 35 \%\) as a proxy for \(\mathrm{F}_{\text {MSY }}\), while \(\mathrm{B}_{\text {trigger }}\) is considered as equivalent to \(\mathrm{B}_{\mathrm{pa}}\).
- Current BRPs:
- \(\mathrm{B}_{\text {lim }}=8075 \mathrm{t}\)
- \(\mathrm{B}_{\mathrm{pa}}=12673 \mathrm{t}\)
- \(B_{\text {loss }}=7507 \mathrm{t}\)
- \(B_{\text {trigger }}=12673 \mathrm{t}\)
- \(\mathrm{F}_{\text {lim }}=\mathrm{NA}\)
- \(\mathrm{F}_{\mathrm{pa}}=\mathrm{NA}\)
- \(\mathrm{F}_{\max }=\mathrm{NA}\)
- \(\mathrm{F}_{0.1}=0.11\)
- \(\mathrm{F}_{\text {MSY }}\) proxy \(=0.13\)

\section*{8. General Comments:}
- First, although unclear, it seems that 2015 is the final year of the assessment. The RG encourages the new index that includes CGFS data past 2015 to be reviewed for use in the assessment. With a different vessel, data from the CGFS from 2015 and later could possibly be used as an additional survey index that covers a different area and time. This will allow the assessment to be updated to 2018. Without running the model until 2018, the current stock status will remain unknown. The RG also recommends continuing efforts for surveying inshore areas and improving the design of the Solent spring and Thames surveys so that they can be used in future assessments. It could be beneficial to expand the survey range for the Solent and Channel trawl surveys, since they only cover a part of the stock range. The RG recommends conducting studies to find if the survey area coverage is still appropriate.
- The WG report is well written and follows the stock annex. The data and data issues are well documented and discussed at length. The WG does an excellent job providing the sources of uncertainty.
- There seem to be a lot of data gaps in at-sea sampling, especially in the French fishery. Because of this, a sensitivity analysis for discard rates is recommended. The RG highly recommends improving on-board sampling coverage and continuing the UK logbook program. A simulation study to find the optimal level of on-board sampling would be helpful.
- The RG recommends continuing and improving the recreational fishery surveys and studies on recreational fishing effort. There seems to be a lot of uncertainty with recreational F. The RG recommends a sensitivity analysis with different levels of recreational \(F\) including a time varying \(F\).
- The RG recommends studies on differences in post-release mortality in the commercial and recreational fleets among sizes of sea bass.
- The RG recommends tagging studies to further understand migration.
- Studies focusing on changes in maturity ogives over time for sea bass are recommended, since there has been no sampling in recent years to determine if the current ogive is still valid.
- As recruitment is variable yet very important to the stock, and temperature has a large effect on recruitment, the RG recommends studies on the effect of temperature on recruitment habitat and the impact of temperature on the stock.
- It would also be useful to run multiple scenarios with changes in recreational F in
the projections. The RG recommends forecasts with changes in selectivity as well, given ICES recommendations for management.
- Model-based abundance indices should be considered. Models, such as a Vector Autoregressive Spatio-Temporal (VAST) model, can be used to model abundance indices and remove the effect of outliers in the survey data.

\section*{9. Technical Comments:}
- In general, some statements need to be reworded. Figures need figure captions, and there are some duplicate figures.
- Mohn's Rho needs to be calculated for the retrospective analysis.
- It is not clear why data from before 1985 are not used.
- The RG would like clarification on if a sensitivity analysis was performed with different natural mortalities. It is not clear from the text. If not, a sensitivity analysis should be performed. On pg. 854, there is a list of M's that were tried in the benchmark (2017), but there are M's that are of higher values in the table on pg. 885 than in the list on pg. 854. So, sensitivity analyses with higher M's should be run as well if they have not been already. Later in the report, it was stated that \(\mathrm{M}=0.24\) in the benchmark, but it was also stated that the assessment assumed \(\mathrm{M}=0.15\). Which M was used in the update?
- On pg. 869 in section 27.2.1, the statement 'fleet 5- other countries plus UK gears not included in fleet 1 ' is not clear. It seems like a lot of UK gears were not included in fleet 1 but then were included in fleets 2 and 3.
- It is not clear which approach is used to estimate recreational catches from 19852014. On pg. 874, an approach is first described but shortly after, the text states 'An average of the two UK effort methods was included'. This is confusing. Additionally, there was an estimate of \(60 t\) of Belgium recreational catches, and it is not clear where this estimate came from.
- The report states that information on ageing precision and calibration will become available after a study in 2015, but this information should be available now, since it is 2019.
- The equation on pg. 891 seems wrong. Wouldn't this result in just the mean abundance in strata s and not the mean abundance?
- The RG would like clarification as to why the first three years of composition data from the CGFS are excluded from use in the model.
- The final estimation year of the model is not clear. The text sometimes refers to the final year being 2016 .
- The maximum age is not clear.
- On pg. 899, the sentence 'A detailed short-term forecast is given in Table 10.1.5.2 assuming that F in 2015 and 2016 is the 2015 values scaled to the average of the previous three years from the assessment.' is not clear. This makes it seem like the projection was carried out in 2015 and 2016. 'Scaled to the average of the previous three years' is also confusing.
- On pg. 899, the report states 'the actual rate of decline in population abundance in recent years is likely to be more uncertain than indicated by the SS3 model confidence limits.' Why is this the case?
- On pg. 899, the report states 'the final package of technical and other
management measures for seabass in 2015, 2016, and 2017 are not fully known at this stage...' This sentence may need to be updated as the management measures from 2015-2017 should be known in 2019.
- It is not clear if the projected landings of 2305 t in 2016 is a result of \(F\) status quo or another F scenario.
- In the references, the citation for Then et al. (2015) is not complete.
- In Table 3, number of landings sampled is not clear. Is this hauls or trips? Additionally, does a blank cell represent 0 or not recorded?
- On pg. 916, the table caption should read 'Sampling of commercial fishery landings by gear...' not 'by area...'. This table format should be improved so that it is easier to read as well.

\section*{10. Conclusions:}
- The assessment of sea bass in divisions 4.b-c, 7.a, and 7.d-h appears well done and describes data and uncertainties in depth. However, the assessment needs to be updated with more recent data (2016-2018).
- The RG suggests the sea bass assessment to be accepted as long as following suggestions are considered: a) update the assessment with more recent data (20152018); b) calculate Mohn's Rho; c) conduct a sensitivity analysis with discard rates; and d) research the effect of temperature on recruitment.

\section*{Cod.27.6a - Cod (Gadus morhua) in Division 6.a (West of Scotland)}
1. Assessment Type: Update
-Last inter-benchmark: 2019
-Last Benchmark: 2012
2. Assessment: Accept with caveats
3. Forecast:
-Short-term: Total biomass (\& SSB) and catch for 2019, 2020, \& 2021 (using Time
Series Analysis (TSA))
-Mid-term: N/A
-Long-term: N/A
4. Assessment Method: TSA
-As outlined in the stock annex
-Data Input (follows benchmark)
- All 5 potential surveys
- Commercial landings data (some data downweighted)
- Discards-at-age data (age compositions only from 1991-2005)
- Recruitment modelled by a Ricker model (except for 1987 because of large 1986 year class)
- Age at full selection \(=6\)

\section*{5. Consistency:}
-All input data for assessment model was data agreed upon in the 2019 inter-benchmark (reflected in the stock annex).
-Advice on stock remains unchanged from 2018 report.
-Methodology of assessment was consistent with the stock annex except for:
- "Additional down-weighting of individual data points to improve TSA assessment model diagnostics (the stock annex acknowledges the need to allow for changes to the variance structures used in the TSA models if they improve model diagnostics)" (Quote from Report).
- "The forecast assumptions differ from those used at previous assessment WGs and those documented in the Stock Annex which have not been discussed or modified since 2008." (Quote from Report).

\section*{6. Stock Status:}
-SSB on decline since 2016.
-Currently at 2,357 t (from TSA).
\(\cdot \operatorname{SSB}\) currently below \(\mathrm{B}_{\text {Trigger }}\left(\mathrm{B}_{\mathrm{pa}} ; 20,000 \mathrm{t}\right)\) and \(\mathrm{B}_{\lim }(14,000 \mathrm{t})\).
-SSB projected for:
o2019: 2,344 t
o2020: 2,013 t
o2021: 2,213 t
-F has been relatively steady since 2014. Currently at 0.668 for 2019 (TSA).

\section*{7. Management and Biological Reference Points:}
- No targeted fishery.
- No HCR found in report or annex (cod management plan discontinued in 2018).
\(\cdot\) TAC ( \(1,735 \mathrm{t})\) is set for bycatch only. Previously, TAC for bycatch of cod was set at \(1.5 \%\) of live weight of total catch from trip.
\(\cdot \mathrm{B}_{\text {trigger }}=20,000 \mathrm{t}\)
- \(\mathrm{B}_{\mathrm{lim}}=14,000 \mathrm{t}\)
- \(\mathrm{B}_{\mathrm{pa}}=20,000 \mathrm{t}\)
- \(\mathrm{F}_{\text {MSY }}=0.29\)
- \(\mathrm{F}_{\text {MSY }}\) Lower \(=0.2\)
- F \(_{\text {MSY }}\) Upper \(=0.41\)
- \(\mathrm{F}_{\mathrm{p} .05}=0.64\)
- \(\mathrm{F}_{\text {lim }}=0.77\)
- \(\mathrm{F}_{\mathrm{pa}}=0.55\)

\section*{8. General Comments:}
-The RG appreciates the WG's efforts to clearly explain their justifications for deviations from the stock annex. The WG's justifications are viewed by the RG as satisfactory. -The RG additionally appreciates the thorough outline of all TSA inputs and outputs.
-The WG discusses an additional methodology by Cook (2019) to initiate conversation concerning drastically different assessment outputs by different methodologies and states that the TSA used for this assessment may not adequately reflect certain aspects of this stock. The RG would appreciate more of a discussion on Cook (2019)'s methods. The RG understands that because of uncertainty in input data that an appropriate assessment type is hard to justify. The RG is not looking for a greater justification of why a TSA was chosen, but simply a larger discussion concerning the uncertainties and what other methodologies show for results.

\section*{9. Technical Comments:}
- Methodology of setting the bycatch TAC for 2019 is not explained. The WG references the landings obligation but does not explain how the TAC was calculated.
-Table 5.16 is too large for the page and so some data cannot be seen.
-Table 5.17 needs further clarification of the abbreviations. The caption is not enough.
-Table 5.18 also needs further clarification of the abbreviations. The caption is not enough. This table also seems to be too large for the page.
-Figure 5.1 has no legend. The color of the grid cannot be used to discern any information.
-Figures 5.3, 5.4, 5.5, \& 5.6 have no axes labels and need more clarification in the captions.
-It is unclear over what time series Figure 5.7 is portraying.
-There is no legend for Figure 5.10.
-Figure 5.11 has no x-axis label.
-Figure 5.12 has no legend.
-CPUE is not explained in terms of units in Figure 5.14.
-Figure 5.15 has no legend.
-Figure 5.18 has no legend.
-Figure 5.19 has no legend and the \(x\)-axis label on the second graph is wrong
-Figure 5.23 has no legend and the \(x\)-axis label on the second graph is wrong.
-Figure 5.25 has no legend and the x -axis label on the second graph is wrong.
-Figure 5.26 has no legend.
-Figure 5.28 has no legend and the x -axis label on the second graph is wrong.
-Figure 5.29 has no legend.
-Figure 5.31 has no legend and the x -axis label on the second graph is wrong.
-Figure 5.32 has no legend and the \(y\)-axis needs a new label.
-Figure 5.38 has no \(x\)-axis label.
-Figure 5.39 has no x -axis labels or y -axis labels.
-Figure 5.40 has no \(y\)-axis labels or legend.
-Figure 5.44 has no x - or y -axis labels or legend.

\section*{10. Conclusions:}
-The RG commends the WG on a well put-together report. The WG followed the stock annex most of the time but provided enough justification when they deviated.
-The RG recommends this report to be accepted under these caveats:
- Address the general and technical comments, especially the third general comment. The RG feels this is an important aspect of the report and is left out of discussion.

\section*{Cod (Gadus morhua) in Division 7.a (Irish Sea)}

\section*{1. Assessment Type: Update}
- Last Benchmark: 2017
- Method changed in 2019
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- No short, medium or long term projections were carried out.
4. Assessment Method: Trend-based assessment using a relative biomass and harvest index

\section*{5. Consistency:}
- A trend-based assessment was used for 2019 instead of using the analytical ASAP model used in 2017 and 2018 because the ASAP model provided a large retrospective bias. As the WG switched to a trend-based analysis this year, the 2019 assessment did not follow the stock annex.
- While the ASAP model showed an increasing trend in SSB since 2010, the biomass index estimated from the trend-based analysis shows no clear trend as these two approaches considered different age groups for SSB.

\section*{6. Stock Status:}
- Biomass index is fluctuating without a clear trend.
- Catches since 2000 have been low and decreasing.
- Landings have been not exceeded the TACs since 2000 except year 2001, 2014 and 2015; the TAC uptake was \(57 \%\) in 2017.
- Discards have been decreasing since 2015.
- The stock status in future years is very uncertain.

\section*{7. Management and Biological Reference Points:}
- Catch advice for 2020 was setting TAC at 516.5 t .
- No reference points available in the assessment as stock has been re-classified as category 3 .

\section*{8. General Comments:}
- The RG considers the report to be very well written. The data and data issues are in general well documented, and moderately discussed.
- The WG added a trend-based analysis in the assessment this year due to large bias of ASAP model. The RG would like to see more comparisons between these two approaches and more discussion on how the new method was chosen over the ASAP model.
- As the assessment did not follow the stock annex, the RG recommends the WG to provide detailed information of the methodology in the report and update the stock annex.
- The RG noted that while ASAP showed an increasing trend in SSB since 2010, the biomass index estimated from the trend-based analysis showed a mostly stable biomass trend. The WG stated that it is because the biomass trend considered only fish from ages 1 to 4 , of
which not all constitute the SSB. The RG recommends the WG to provide more evidence to support their selection of age range.
- More discussion of sources of uncertainty and quality of data, as well as further discussion of the alternative assessment model would strengthen this update.

\section*{9. Technical Comments:}
- Section 7.1. "The landings in 2018 were observed at \(235.9 \mathrm{t}, 214.9 \mathrm{t}\) after re-allocation of \(20 t\) of Irish landings and despite the TAC of \(695 t\) only increased slightly (Table 7.1)." The RG noted that the observed landings is 234.9 according to the Table 7.1.
- Table 7.2. - The unit of landings and TAC need to be added.
- Table 7.8. - Data for 2014 should be added or provide an explanation of why they are not available.
- Figure 7.1. - The whole plots should be well within the bounding box.
- Figure 7.2. - The axes should be labeled.
- Figure 7.3. - The figure will be clearer if the top block showing "Raw stock weights" is removed.
- Figure 7.4. - "Log ratio of ages in commercial catches." The description seems to be inconsistent with the figure.
- Figure 7.9. "Commercial fleet catch-at-age residuals." The RG is not clear with which figures are described by the caption. The RG suggests numbering the figures and then classifying the number in the caption.
- Figure 7.15. - The text on the plot is hard to read.
- Figure 7.7. - The legend to the right is not clear and needs to be clarified.
- Section 7.1. In the section of "Management applicable to 2018", the WG stated that "Technical regulations in force in the Irish Sea, ... are described in Section 7.2 and 7.10". The RG noted that there is no such information available in Section 7.2, and there is no Section 7.10 in the report. The RG recommends the WG to revise and improve the content.
- Section 7.2. In the section of "Survey data used in assessment", the WG stated that the age range of NIGFS-WIBTS-Q1 and UK-FSPw is 1-4 and 2-6, respectively. However, the stock annex states those age ranges are both 1-5. The RG recommends the WG to explain the reason for the inconsistency.
- Section 7.2. The WG stated that "NIGFS-Q1 was used as a biomass trend, multiplying the relative abundance at age/nautical miles by the weight at age and applying the two over three rule subsequently." The RG recommends the WG to give more information to explain this clearer, such as a description of two over three rules.

\section*{10. Conclusions:}
- The assessment of cod in Division 27.7.a. appears well done. However, some grammar and sentence structure errors can make interpretation of the output a bit challenging.
- The WG decided to use a trend-based analysis for assessment this year instead of using the conventional ASAP model due to its large retrospective bias. The RG suggests: a) provide more information about the retrospective bias; b) include complete methodology of the new method in the assessment report or update the stock annex; and c) conduct more discussion on model comparisons.

\section*{Cod (Gadus morhua) in Divisions 7.e-k (Eastern English Channel and southern Celtic Seas)}
1. Assessment Type: Update
- Benchmarked in 2012
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- No short-term projection was performed.
- No medium or long term were carried out.
4. Assessment Method: Extended Survivor Analysis (XSA)

\section*{5. Consistency:}
- To ensure the consistency of data processing at the international level, the same rules were applied each year for the allocation procedure for TAC: fill unsampled strata using as much as possible the same metier and quarter, regardless of area and country.
- Biological parameters remain unchanged since the 2012 WKROUND benchmark.
- The recent technical measures introduced in the Celtic Sea (square mesh panels) were not expected to significantly reduce catches of Celtic Sea cod or improve the selection pattern.
- Recent recruitment was low and the short-term outlook was very dependent on recruitment for this stock.
- The final assessment was run with the same settings as established by WKROUND 2012. Discards were not included in the assessment.
- This year's model output suggested a substantial downward revision in SSB and recruitment in recent years and substantial upward revision in F.
- The advice remains the same, because the short-term outlook for the stock remains unchanged.

\section*{6. Stock Status:}
- The current SSB is below \(\mathrm{B}_{\text {trigger }}\) and \(\mathrm{B}_{\mathrm{pa}}\). SSB has decreased from 12,492 t in 2012 to \(1,783 \mathrm{t}\) in 2018.
- In 2018, F is estimated at 0.83 which is above \(\mathrm{F}_{\text {lim }}, \mathrm{F}_{\mathrm{pa}}\) and well above \(\mathrm{F}_{\text {MSY }}\).
- F was above \(\mathrm{F}_{\text {MSY }}\) during the 1971-2018 period but decreased between 2000 and 2010 and increased again after. F fluctuated around \(\mathrm{F}_{\text {lim }}\) in recent years.
- Catches have been around 5,000 t since 2000. Total catch in 2018 was \(1,385 \mathrm{t}\).
- SSB is below MSY \(B_{\text {trigger }}\) and \(B_{p a}\) since 2000, except for SSB in 2012 as a consequence of a very good recruitment year. Since 2004, SSB has been below \(\mathrm{B}_{\mathrm{lim}}\), except during the 2011-2013 period.
- Recruitment has been highly variable over time with occasional very high recruitment followed by a period of low recruitments. Since 2012, recruitment has been very weak except for the 2014 year class, which is above average. Recruitment estimated in 2017 and 2018 were remarkably low.

\section*{7. Management and Biological Reference Points:}
- No management plan has currently been established for cod in Subareas 7e-k.
- Technical measures applied to this stock were a minimum mesh size (MMS) for beam and otter trawlers in Subarea 7 and a minimum landing size (MLS) of 35 cm.
- BRPs have been estimated using the agreed ICES guidelines in 2016.
- \(B_{l i m}\) and \(B_{p a}\) remained unchanged.
- \(\mathrm{F}_{\text {lim }}\) was defined according to segmented regression with \(\mathrm{B}_{\mathrm{lim}}, \mathrm{F}_{\mathrm{pa}}\) was equal to \(\mathrm{F}_{\text {lim }} / 1.4\).
- A TAC of \(1,610 \mathrm{t}\) was set for 2019.
- \(\mathrm{B}_{\text {trigger }}\) is considered as equivalent to \(\mathrm{B}_{\mathrm{pa}}\).
- Current BRPs:
- \(\mathrm{B}_{\lim }=7,300 \mathrm{t}\)
- \(\mathrm{B}_{\mathrm{pa}}=10,300 \mathrm{t}\)
- \(\mathrm{F}_{\text {lim }}=0.8\)
- \(\mathrm{F}_{\mathrm{pa}}=0.58\)
- \(\mathrm{F}_{\mathrm{MSY}}=0.35\)
- \(B_{\text {trigger }}=10,300 \mathrm{t}\)

\section*{8. General Comments:}
- The WG report is well written and follows the stock annex. The data and data issues are well documented and discussed. The WG did extensive work for this year's assessment.
- WG added up the quarterly data of catch numbers-at-age and catch weights-at-age for their landings of France, Ireland, and the UK (E+W) and raised to international landings with consideration of Belgian data. However, the WG mentioned that there is evidence that misreporting has increased from 2002 when quotas became restrictive with a maximum in 2008. Although misreporting has decreased since then, it still leads to certain uncertainty in the assessment results, because XSA assumes high quality catch-at-age and weight-at-age data for every time step.
- Discard estimates were available for countries; however, the assumption and estimation methods were different. For example, only data from the at-sea observer program were included in this report for France, and for Ireland, the assumption was that the discards are mainly at-age 1 , and the estimates are very uncertain. For fleets from Belgium, the modal distribution of discards was around 30 cm . The RG believes that consistency analysis for discard estimation is necessary.
- In 2017, the French EVHOE survey was not conducted due to technical difficulties at the beginning of the survey. The IR-FR combined tuning index used in the assessment is only composed of Irish data for 2017. This change of survey design and data will introduce uncertainty and bias in the assessment and forecast. The RG suggests setting different levels of catchability to conduct a necessary sensitivity and uncertainty analysis.
- This year, no changes were made to the input data, and two ongoing surveys, both part of the DCF, IBTS Q4 (FR-EVHOE \& IR-GFS7gj combined) were used to assess this stock. The survey index was a combined index based on both French IR-GFS and FR-Evhoe Q4 data. In addition, the WG evaluated the influence of having more symmetrical grid cells in data aggregation process through the comparison of new grid ( 0.25 deg lat \(x 0.5\) deg long) and the historic grid ( 0.5 deg lat \(x 0.5\) deg long). The RG suggests more detailed description of the comparison results.
- Natural mortality was assumed to be constant for the whole range of years and was age dependent. The RG suggests adding more biological clarification to justify the choice of fixed natural mortality, coupled with a structured sensitivity analysis to evaluate impacts of uncertainty in natural mortality.
- The initial time series of input data for retrospective analysis was 1971-2013 in the report. Due to the long time series of input data, the RG recommends removing more years of data to increase the number of retrospective peels to better reflect the performance of the methods.
- The WG indicated that a rescaling in SSB, recruitment and F is evident by comparing this year's assessment results with last year's results and by analyzing the retrospective analysis. The RG suggests a more detailed description of model performance according to retrospective analysis.

\section*{9. Technical Comments:}
- No detailed description about the selectivity pattern was in the report.
- Page 7 Surveys and commercial tuning fleet - The RG believes that the standardization process should be clarified with more technical details and references. The purpose of standardization should also be clarified.
- Page 7- "The comparison of runs with and without tuning indices indicates that the majority of the information comes from the catch-at-age matrix (Figure 8.12)". Replace Figure 8.12 with Figure 8.11b.
- Page 8-"Length distributions of 2017 discards provided by countries for sampled strata and quarter are shown Figure 8.8a-d". Replace Figure 8.8a-d by Figure 8.3a-d.
- Figure 8.2.a. - "2018" is an incorrect abscissa scale.
- Table 8.3 - There should be a unit for the catch (landings) weight at age.
- Figure \(8.3 \mathrm{a}-\mathrm{d}\) - The landings and discards length distributions should be separate. Because the magnitudes of these are quite different, the combined figure will lead to unclear distribution of landings length distribution.
- Table 8.4 - There should be a unit for the stock weight at age.
- Figure 8.5 a - Fleet information is needed.
- Figure 8.6 b - The label of the y -axis is incomplete.
- Table 8.6 - The format of the table should be reorganized.
- Figure 8.7 - The significance of bubbles needs to be defined.
- Table 8.7 - The format of the table should be reorganized.
- Table 8.8 - There should be a unit for the number-at-age.
- Figure 8.8 - The average F-at-age scaled by \(\mathrm{F}_{\text {bar }}(2-5)\) was used, the RG recommends providing the reasons for this processing.
- Table 8.9 - There should be a unit for the number-at-age.
- Table 8.10a - There should be units for recruitment, SSB, catch, landings and TSB.
- Figure 8.11a - Types of input data for the assessment should be made clear earlier in the report.
- Figure 8.11a - The RG suggests that the caption should clearly indicate number of years considered in the retrospective analysis.
- Figure 8.11b - The color of lines representing runs with and without tuning indices are not obviously different.
- Stock annex: page 8 "Figure 3 show that the correlation between the indices is very high, which further validates the use of the new index in the assessment. Details presentations of both estimates are presented in Table A" - Replace Figure 3 by Figure C.

\section*{10. Conclusions:}
- The assessment of Cod (Gadus morhua) in divisions 7.e-k (Eastern English Channel and southern Celtic Seas) appears well done and XSA residuals and diagnostics do not highlight any problems regarding the input data and model fit in the report. However, due to the uncertainty in catch-at-age and weight-at-age data, the RG suggests running other assessment models for comparison.
- Survey design and data were changed in 2017, which will introduce uncertainty and bias in the assessment and forecast. The RG suggests setting different levels of catchability to conduct a necessary sensitivity and uncertainty analysis.
- Given the strong retrospective pattern of this year assessment, the RG also suggests a more detailed description of model performance according to the retrospective analysis. In addition, the RG recommends removing more years of data to increase the number of retrospective peels.
- Although the WG mentioned that the non-inclusion of undersized discards in the assessment might cause retrospectives bias, more impact issues should be considered to illustrate the retrospective bias. For example, there was no description on whether the life history of cod has changed significantly during time series, which may result in strong retrospective patterns. In addition, the effect of environmental variables on cod should also be taken into consideration.
- The RG suggests cod (Gadus morhua) in divisions 7.e-k be accepted as long as the above concerns are addressed.

\section*{had.27.7b-k: Haddock in divisions 7.b,c,e-k}
1. Assessment Type: Update
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- Short term projection - Carried out by Age Structured Assessment Program (ASAP) from 2019 to 2021.
- No medium or long term projections were carried out.
4. Assessment Method: ASAP was used. XSA was also used for quality control purposes.

\section*{5. Consistency:}
- The combined French/Irish survey has nearly full spatial coverage of the assessment area. The survey has good internal consistency.
- Additional technical measures have been introduced to reduce the high levels of discards recently observed in the Celtic Seas.
- The estimate for SSB is biased in the most recent year.
- Recruitment was large in 2018 due to the large recruitment observed in the surveys.
- F has been overestimated and revised downwards since 2008 in the assessment due to the strong 2009 cohort causing a conflict with the catch data.

\section*{6. Stock status}
- The SSB peaked in 2011 as the very strong 2009 year class matured; this cohort was followed by three years of below-average recruitment which led to a rapid decline in SSB after 2011. SSB has since stabilized.
- Recent recruitment has varied around the average, with a notable peak in 2009 and in 2018.
- F has been above \(\mathrm{F}_{\text {mSy }}\) for the entire time-series.
- Although F is consistently over \(\mathrm{F}_{\text {MSY }}\), the stock size has not decreased, suggesting the stock is robust to overfishing. If recruitment were to be consistently low, then SSB could decline below \(\mathrm{B}_{\text {trigger }}\).
- Total landings in 2018 were 6590 t.
- Landings decreased from 6685 t in 2017.

\section*{7. Management and Biological Reference Points:}
- The TAC in 2018 was 8329 t .
- MSY and PA reference points were defined in 2016 and have not been altered.
- The minimum size is 30 cm .
- MSY advice currently uses \(\mathrm{F}_{\text {MSY }}\) (median point estimates of EqSim with a segmented regression stock-recruitment relationship), while \(B_{\text {trigger }}\) is considered as equivalent to \(B_{p a}\).
- Current BRPs:
- \(B_{\lim }=6700 t\)
- \(\mathrm{B}_{\mathrm{pa}}=10000 \mathrm{t}\)
- \(\mathrm{B}_{\text {trigger }}=10000 \mathrm{t}\)
- \(\mathrm{F}_{\text {lim }}=1.41\)
- \(\mathrm{F}_{\mathrm{pa}}=0.89\)
- \(\mathrm{F}_{\max }=\mathrm{NA}\)
- \(\quad \mathrm{F}_{\mathrm{MSY}}=0.40\)

\section*{8. General comments:}
- The WG report is well written and organized. The data and data issues are well documented and discussed at length. The WG used two different models to further understand the robustness of the results, which the RG commends.
- The retrospective patterns are concerning. The RG recommends conducting a retrospective analysis with the XSA model and comparing the retrospective bias between the XSA and ASAP models.
- The RG recommends research on variables, such as environmental conditions, that drive recruitment. Perhaps environmental variables should be incorporated into the assessment.
- Since recruitment is variable for this stock, it would be beneficial to reconsider recruitment configurations in the short-term projections or try scenarios with high and low recruitment values.

\section*{9. Technical comments:}
- The RG would like more clarification as to why this stock is still suitable for category 1 although there are large retrospective patterns.
- The RG would like clarification on why recruitment in the short-term projection a geometric mean of recruitment from 1993-2016 and not until 2018.
- On pg. 6, the report states 'The results have been published earlier this year (ICES, 2016b)...'. This is an old citation, so this sentence needs to be updated.

\section*{10. Conclusions:}
- The assessment of haddock in divisions 7.b-k appears well done and organized. The RG appreciates the use of two models for quality control purposes.
- Given the two model outputs are similar, the RG suggests the haddock assessment to be accepted as long as the following suggestions are considered: a) Compare the retrospective bias between XSA and ASAP models; b) research environmental variables that drive recruitment; and c) reconsider recruitment configurations in the projections (i.e. scenarios with high and low recruitment values).

\section*{lez.27.4a6a: Megrim (Lepidorhombus ssp.) in divisions 4.a and 6.a (northern North Sea, West of Scotland)}
1. Assessment Type: Update
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- Accept with caveats
- Short-term projections - Carried out according to the methods described in the stock annex.
- No medium or long term were carried out.
4. Assessment Method: Bayesian state-space surplus production model.

\section*{5. Consistency:}
- The model configuration is the same as the 2018 assessment. The 2019 assessment is only an update of 2018 assessment with new catch and survey data.
- The 2019 assessment compares the estimates of key parameters with that of assessments from previous years. The estimated K, MSY, \(\mathrm{B}_{\text {MSY }}, \mathrm{B}_{\text {lim }}\), and \(\mathrm{B}_{\text {trigger }}\) in 2019 are lower than their estimates in 2018. The estimated \(r\) and \(\mathrm{F}_{\text {MSY }}\) in 2019 are similar but slightly less than the estimates in 2018.
- The model outputs from the 2019 assessment suggest a downward revision of the biomass and a stable trend of fishing mortality compared with that of 2018.
- The 2019 assessment presents the results of retrospective analysis, which was not conducted in 2018. There is no retrospective pattern detected and the magnitude of the Mohn's Rho values is less than 0.1.
- The status of the stock has not changed since last year.

\section*{6. Stock Status:}
- Biomass has consistently been above \(\mathrm{B}_{\mathrm{MSY}}\) and shows an increasing trend since 2005. The estimated biomass in 2018 is \(37,062 \mathrm{t}\) and the \(\mathrm{B}_{2018} / \mathrm{B}_{\mathrm{MSY}}\) is 1.681 .
\(\bullet\) F has consistently been below \(\mathrm{F}_{\text {MSY }}\) and shows a declining trend since the late 1990s. The estimated \(\mathrm{F}_{2018}\) is 0.08 and \(\mathrm{F}_{2018} / \mathrm{F}_{\text {MSY }}\) is 0.40 .
- Total catch in 2018 was \(3,258 \mathrm{t}\). The ICES estimated landings without discards in 2018 were \(3,003 \mathrm{t}\), which is well below the TAC \((7,958 \mathrm{t})\).

\section*{7. Management and Biological Reference Points:}
- ICES advised that when the EU multiannual plan is applied, catches in 2020 that correspond to the F ranges in the plan should be between 6450 t and 8350 t .
- \(\mathrm{F}_{\text {MSY }}, \mathrm{B}_{\text {MSY }}\), and MSY were directly estimated in the model and the values varied when new catch and survey data are added into the assessment.
\(\bullet B_{p a}\) and \(B_{\text {lim }}\) were defined as \(50 \% \times B_{\text {MSY }}\) and \(30 \% \times B_{\text {MSY }}\) respectively from the previous assessment and remained unchanged.
- \(\mathrm{F}_{\text {lim }}\) is the F brings the stock to \(\mathrm{B}_{\text {lim }}\) and is defined as \(1.7 \times \mathrm{F}_{\text {MSY }}\) from the previous assessment. It is remained unchanged in the 2019 assessment.
- MSY \(\mathrm{B}_{\text {trigger }}\) is set to \(\mathrm{B}_{\text {MSy }}\) because the stock has been fished below \(\mathrm{F}_{\text {MSY }}\) for more than 10 years.
- Current BRPs:
- MSY \(=5123 \mathrm{t}\)
- \(\mathrm{F}_{\mathrm{MSY}}=0.25\)
- \(\mathrm{B}_{\mathrm{MSY}}=22058 \mathrm{t}\)
- \(\mathrm{B}_{\lim }=6617 \mathrm{t}\)
- \(\mathrm{B}_{\text {trigger }}=11029 \mathrm{t}\)

\section*{8. General Comments:}
- The WG report is well written and the stock assessment follows the methods that are described in the stock annex. Both fishery-independent and fishery-dependent data are well documented. In addition, the WG does an excellent job discussing the sources of uncertainties that are associated with the data and providing recommendations for next benchmark.
- The RG recommends addressing several identified data uncertainties through a structured sensitivity analysis.
oThe WG documented that there are a few discrepancies with the two types of landing estimates. Because the current assessment only uses estimated landings from ICES and the discrepancies between the two types of landing estimates in Subarea 6.a. were relatively large before 2000, the RG recommends conducting an exploratory run using the official landings and quantifying the influence of different landing inputs on the assessment results. The RG also recommends providing explanations on how to adjust the official landings to InterCatch/ICES landings, especially when the estimated ICES landings in Subarea 4 and 2.a. were less than official landings reported to ICES.
- The discards data were available since 2013, and the WG assumed a linear decline in discards from \(30 \%\) to \(15 \%\) over time from 1985 to 2012 in the assessment. The RG recommends providing justification on such a strong and specific assumption. The RG also recommends providing the estimates from the scenario of a fixed \(15 \%\) discard proportion over the full landings, which is described as another assumption in the stock annex, as a sensitivity run. In addition, the RG recommends considering discards with a random-walk pattern since there is no documented supporting evidence on the trend of discards before 2013.
\(\circ\) As both landings and discards were uncertain, the total catch of landings and discards could be modeled by a lognormal distribution with mean equal to the current catch input and variance corresponding to a constant coefficient of variation (Marandel et al. 2016). The RG recommends developing this scenario as a sensitivity run.
- As two AMISS surveys collect data in the same season and same subregion (i.e. 6.a), the RG supports the recommendation from the WG on combining the information from the two surveys to develop one continuous index for the future benchmark assessment. However, the correlation between the indices of the two surveys needs to be carefully examined before combining the information. The depth ranges of the two surveys are different, and one of the surveys is not designed for monitoring megrim. The methodology of calculating the abundance index and difference in catchability of the two surveys needs to be explained.
- The WG uses a Bayesian state-space surplus production model with the assumption that the surplus production is maximized at \(50 \%\) of unfished biomass. The RG recommends conducting the assessment with a more generalized approach by incorporating a shape parameter to control the level of biomass as a proportion of unfished biomass at which surplus production is maximized (i.e. Pella-Tomlinson model).
- The WG summarizes the results of the final run in various tables and figures. The RG recommends the WG conduct diagnostics to check the performance of the final run and interpret the results in the report.
- If the WG conducted more than one run, the RG recommends providing values of DIC and mean error between predicted and observed indices for comparing results from multiple model runs.
- The RG recommends the WG describe the posterior distributions in figure 13.10 and providing Gelman and Rubin statistics to verify the convergence of the model run.
- The RG also notices the temporal pattern in the residuals from figures 13.8, 13.9, and 13.11. The model does not fit the two survey indices well. The RG recommends examining the issues from multiple aspects. Because the GLM model-based abundance indices from SCO Vla Q4 and SCO IVa Q3 already showed a trend of being larger than the arithmetic mean of the abundance indices, the RG recommends exploring a GAM to derive model-based abundance and check the performance of the model using cross validation. In addition, the WG could check the correlation among these abundance indices and assign different indices with different weights in the stock assessment.
- The WG conducted the retrospective analysis and the results indicate no strong retrospective pattern. However, given the maximum age of the species is about 16 years, the RG recommends removing more years to cover at least the lifespan of the species.
- The WG provides results of short-term projections under various scenarios. It is not clear to the RG that whether the projections are stochastic or deterministic. The RG recommends using WinBUGs to conduct stochastic projections and plotting the projections of biomass from 2019 to 2021 with the median value and \(95 \%\) confidence intervals.

\section*{9. Technical Comments:}
- The WG uses different terms to describe the stock assessment model in the report. Both "Schaefar Surplus production process model" and "Bayesian state-space production model" appeared in different places. The RG recommends using the term "Bayesian statespace surplus production model" and consistently use this term in the report.
- The RG recommends moving all tables and figures in the main text to the Tables and Figures section after the references and providing captions.
- The WG uses InterCatch in the text to describe the catch estimates but uses ICES landings for catch estimates in the table 13.2 and 13.3. The RG recommends clarifying the differences between the two terms.
- The WG used two-stage GLM to derive abundance indices from various surveys because the IBTS surveys exhibit a large proportion of zeros. The RG recommends exploring other families such as Tweedie to account for the large proportion of zeros in the data and other model such as a GAM for comparison.
- The WG mentioned "SSB" under the section of uncertainties and bias in assessment and forecast. The RG recommends changing "SSB" to "biomass" because there is no clear definition of SSB in the report.
- Change "cpue" to uppercase "CPUE".
- For the table under the section 13.5-MSY reference points, please explain the superscript b) and d).
- Table 13.4: Change "surplus production model" to "Bayesian state-space surplus production model".
- Table 13.6: Clarify whether the estimated values are mean or median in the caption. Provide values for \(\mathrm{B}_{\mathrm{pa}}, \mathrm{F}_{\text {lim }}\), MSY \(\mathrm{F}_{\text {lower }}\), and MSY \(\mathrm{F}_{\text {upper }}\) as well.
- Table 13.7: Clarify whether the estimated values are mean or median. Do the "high" and "low" represent \(95 \%\) CI of the estimates? Provide figures of estimated B/BMSY and F/FMSY over time with \(95 \%\) CI.
-Table 13.8: The RG suggests using the average values from 2016 to 2018 for the assumption of discard rate.
- Figure 13.1: Change "2017" to "2018" in the caption. Provide number of stations per year in each figure.
- Figure 13.2-13.5: Add the meanings of " + " and gray dot in the legend.
- Figure 13.6: Plot 95\% CI of the abundance indices.
- Figure 13.8: Plot 95\% CI of the Delta-Gamma mean indices. Change the range of y axis for the figure on the top left.
- Figure 13.9: Expand the range of the \(y\) axis for several figures. Clarify the scale of the values, are they on log scale?
- Figure 13.10: Explain the index number in the title of each figure and which survey index they are representing in the caption.
- Figure 13.14: Use different symbols to highlight the status in initial year and terminal year.

\section*{10. Conclusions:}
- The assessment of megrim in divisions 4.a and 6.a appears well-done. Proper methods were used following the stock annex. The results indicate no large retrospective errors. The RG recommends this stock assessment to be accepted with caveats.
- Given the predicted survey indices in 4. a show that the model did not match the observed data well and there are a few potential issues with input data, the RG suggests megrim to be accepted as long as following concerns are addressed: a) conduct an exploratory run to capture the uncertainty of total catch by modeling catch with variance corresponding to a constant coefficient of variation; b) explore a GAM to derive model-based abundance indices, and assign different indices with different weights in the stock assessment; and c) conduct the assessment with a more generalized approach by incorporating a shape parameter to control the level of biomass as a proportion of unfished biomass at which surplus production is maximized.

\section*{Plaice (Pleuronectes platessa) in division 27.7.a (Irish Sea)}

\section*{1. Assessment Type: Update}
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- Short term projection (2 years) - Implementing the management plan for this stock with \(\mathrm{F}_{\text {MSY }}=0.196\) leads to a total catch of 3299 t in 2020 and SSB of 18354 t in 2021. - No medium or long term forecast was carried out.
4. Assessment Method: Age-based analytical assessment using landings and discards (implemented in State-space Assessment Model (SAM))

\section*{5. Consistency:}
- The Aarts and Poos (AP) model was replaced by the state-space assessment model (SAM) in 2017.
- The estimated selectivities are split into the landed and discarded components. Until the early 1990s, the fleet selectivity had the highest values for fish at age 4. This selectivity shifted to age 5 in the late 90 s and early 2000s. Since the late 2000s, landings gradually fell over time to very low values relative to discards, which became dominant and expanded to the older aged fish during the most recent years.
- Recruitment is fluctuating without an overall trend, and it is estimated at its lowest values in 2017 and 2018.
- A Mohn's rho analysis for a 5-year peel resulted in values of \(0.23 \%\) for recruitment, \(4.64 \%\) for SSB and \(-5.47 \%\) for \(\mathrm{F}_{\mathrm{bar}}\).
- The assessment methodology provided is robust. A serious retrospective pattern did not appear.

\section*{6. Stock Status:}
- The stock is within safe biological limits. Fishing pressure was at a low level, and SSB was at a high level.
- F declined from high levels in the 1980s to very low levels in the early 1990s and has been less than 0.1 since 2013.
- F estimated for 2016-2018 has fluctuated around similar values, from 0.050 (2016) to 0.066 (2017) and 0.064 (2018).
- Catch has decreased to low levels and since 2006. Most of the catch was discarded.
- SSB increased reaching the highest value in 2016 and decreased in 2017.
- Estimated recruitments are highly variable. There was an increasing trend in recruitment before 2015. It dropped to the lowest values in 2017 and 2018.

\section*{7. Management and Biological Reference Points:}
- Commission regulations included a prohibition on the use of demersal trawls, enmeshing nets or lines within the main cod spawning area in the northwest Irish Sea between the 14th of February and 30th of April.
- Some derogations were permitted for Nephrops trawls and beam trawlers targeting flatfish.
- There is a Minimum Conservation Reference Size (MCRS) of 27 cm .
- Plaice is managed by TAC in 2018 (1793 t) and 2019 ( 3075 t).
- There is a mismatch between the minimum landing size and the mesh size of the gear being used.
- Implementing the management plan for this stock with F MSY leads to a total catch of 3299 t (1931 t of landings and 1368 t of discards (including dead and survivors)) in 2020 and a SSB of 18354 t in 2021.
- Current BRPs:
- \(\mathrm{B}_{\mathrm{lim}}=3958 \mathrm{t}\)
- \(\mathrm{B}_{\mathrm{pa}}=5294 \mathrm{t}\)
- \(\mathrm{F}_{\text {lim }}=0.495\)
- \(\mathrm{F}_{\mathrm{pa}}=0.355\)
- \(\mathrm{F}_{\mathrm{msy}}=0.196\)

\section*{8. General Comments:}
- The data and data issues are well documented and discussed at length. The data were abundant to support the SAM model to assess the plaice stock. The WG did extensive work for this year's assessment including running other assessment models for comparison.
- The RG encourages the WG to update the stock annex to provide information needed to understand the assessment.
- The WG analyzed diagnostic outputs regarding residuals of catch and survey data and retrospective patterns for years 2007-2017. The results suggest the methodology seems to be robust without a serious retrospective pattern but the RG is still concerned about the model performance and reference points (e.g., \(\mathrm{F}_{\text {msy }}\) ). Plaice biomass surveyed in 2019 (NIGFS-WIBTS-Q1) is lower than that in 2018 (Table 21.3a). Plaice biomass surveyed in NIGFS- WIBTS-Q4 and UK(E\&amp;W)-BTS-Q3 had a downward trend since 2014 (Table 21.3b and 21.4). Therefore, the total biomass of plaice in the division 27.7.a may decline in 2019. The SAM model predicted the value of \(\mathrm{F}_{\text {msy }}\) to be 0.196 for 2020 which was higher than \(\mathrm{F}_{\mathrm{msy}}(0.169)\) recommended for 2019. The RG recommends the WG to evaluate whether the BRPs are appropriate.
- Focusing on the most recent two years (2017 and 2018), the projections with the SAM model in 2019 are different from those in previous assessments. The SAM model (2017)
predicted the SSB to be 21356 t in 2017 (ICES, 2017). This value was much higher than that from the SAM model in 2019 (Low: 11448; Mid: 15622; High: 21319). The SAM model (2018) predicted the SSB was 22077 t in 2018 (ICES, 2018). This value was close to 24422 (Low: 12572; Mid: 17522; High: 24422).
- The minimum Conservation Reference Size ( 27 cm ) has been implemented for many years. As the model is age-based, is this management considered in the model? If so, please supplement relevant content in the report or annex.
- The data of discarded and retained catches are length-based (Fig. 21.4 and 21.5). As the model is age-based, how did the WG use them to estimate the discards at age? If using the von Bertalanffy growth model or other formulas to transform the data, please supplement formulas and parameters in the annex.
- The negative values of catch residuals are apparent in ages 8+ from 1998 on. A pattern of negative residuals between 2004 and 2009 is presented in the residuals of the NIGFSWIBTS due to large fluctuations in the SSB indices. The RG recommend the WG to evaluate the causes of these residual patterns.
- As natural mortality is usually not invariant, the RG encourages the WG to evaluate the model performance when incorporating stochastic effects into natural mortality. A recent article (Aldrin et al., 2019) may be of interest to the WG.

\section*{9. Technical Comments:}
- In some fish stock assessment models, age-structured observations are assumed to be statistically independent. The study of Berg and Nielsen (2016) suggested catch-at-age data should not be assumed independent. If correlation is not considered in this assessment, the RG encourages the WG to consider correlation in the model used for stock assessment.
- The report mentioned that "A total of 1000 samples are generated from the estimated distribution of survivors". What does the estimated distribution of survivors refer to?
- Table 21.13 - Please clarify the definition of "Low", "Mid", and "High".
- Figure 21.1, 21.2a, 21.2b, 21.7, 21.10, 21.11, 21.12 - Please add x -axis labels as "Year".
- Figure 21.13 and 21.14 - Please add the \(y\)-axis labels.

\section*{10. Conclusions:}
- The assessment of plaice is well done and trains the model with updated observations such as discard data.
- Given the model shows an adequate fit to the data, the model can be used to assess the plaice stock. The methodology seems to be robust without a serious retrospective pattern. Based on the comments provided by the RG, the report can be accepted after the following concerns are addressed: a) update the stock annex so that the reader can better
understand the model; b) reevaluate some reference points (e.g., \(\mathrm{F}_{\text {msy }}\) ); and c) evaluate impacts of stochastic effect of natural mortality on the model performance.

\section*{27 7.e: Plaice (Pleuronectes platessa) in the Western English Channel}

\section*{1. Assessment Type: Update}

\section*{2. Assessment: Accept}

\section*{3. Forecast:}
- Plaice in 7.e continues to be treated as a category 3.2.0 stock and the assessment is indicative of trends only.
- No short, medium or long term forecasts were carried out.

\section*{4. Assessment Method:}
- An XSA based on landings data only was used for the assessment.
- As this was an update assessment, full data screening, tuning data and extensive exploratory XSA trials were not carried out.

\section*{5. Consistency:}
- The assessment shows a high consistency with last year's assessment. Relative values presented for recruitment, SSB, and F estimates had similar temporal trends to those presented in previous assessments.
- The Mohn's rho values for this assessment are very low and well below the threshold of \(20 \%\) imposed by ICES for 2019 assessments (i.e. the current assessment indicates a very high consistency).
- The assessment contains a certain degree of uncertainty due to excluding discards and is likely to be overly optimistic. F is likely to be higher, and SSB lower than estimated by the current assessment.

\section*{6. Stock Status:}
- F surpassed \(\mathrm{F}_{\text {MSY }}\) in 2016 and has been above since then. The SSB is well above \(\mathrm{B}_{\text {trigger }}\), \(\mathrm{B}_{\mathrm{pa}}\) and \(\mathrm{B}_{\mathrm{lim}}\).
- A combination of above average recruitment and a reduction in F has increased SSB since 2008 to reach the highest level on record in 2016. However, since then, the SSB has decreased but is still a high level.
- F gradually increased from the 1980s up until the 2000s, peaking briefly in 2007. Following a large reduction in F in 2009, this assessment shows a general decline that has reached the lowest levels on record in 2015. Since then, F has increased again.
- This assessment estimates that recruitment has been above the long-term geometric mean (1980-2018) between 2010 and 2015 and below afterward. However, the optimistic stock development in recent years is uncertain due to assessment uncertainty and omitting
discard information. The decision to omit discard data is mainly due to uncertainty in the actual discard rate and an unknown proportion of surviving plaice in the discards.

\section*{7. Management and Biological Reference Points:}
- There is no management plan in place for this stock.
- The current reference points are the ones calculated at WKMSYREF4 (ICES, 2016a) and shown in the following table:
\begin{tabular}{|c|c|c|c|c|}
\hline Framework & Reference point & Value & Technical basis & Source \\
\hline \multirow[t]{2}{*}{MSY approach} & \[
\begin{aligned}
& \text { MSY } \\
& \text { B }_{\text {trigger }}
\end{aligned}
\] & 2443 t & \(\mathrm{B}_{\text {pa }}\) & \begin{tabular}{l}
ICES \\
(2016a)
\end{tabular} \\
\hline & \(\mathrm{F}_{\text {MSY }}\) & 0.238 & Eqsim run with segmented regression with breakpoint at \(\mathrm{B}_{\text {loss. }}\). FmSy was taken as the peak of the median landings yield curve. & \begin{tabular}{l}
ICES \\
(2016a)
\end{tabular} \\
\hline \multirow[t]{4}{*}{Precautionary approach} & B \({ }_{\text {lim }}\) & 1745 t & Bloss & \begin{tabular}{l}
ICES \\
(2016a)
\end{tabular} \\
\hline & \(\mathrm{B}_{\mathrm{pa}}\) & 2443 t & \(1.4 * \mathrm{~B}_{\mathrm{lim}}\) & \begin{tabular}{l}
ICES \\
(2016a)
\end{tabular} \\
\hline & \(\mathrm{F}_{\mathrm{lim}}\) & 0.88 & Based on segmented regression simulation of recruitment without error & \begin{tabular}{l}
ICES \\
(2016a)
\end{tabular} \\
\hline & \(\mathrm{F}_{\mathrm{pa}}\) & 0.63 & Flim* \(\exp (-1.645 * \sigma) ; ~ \sigma=0.2\) & ICES \\
\hline
\end{tabular}

\footnotetext{
* The value for MSY \(\mathrm{B}_{\text {trigger }}\) is not the value published in WKMSYREF4. The advice drafting group in 2017 and 2018 decided to base MSY \(\mathrm{B}_{\text {trigger }}\) on \(\mathrm{B}_{\mathrm{pa}}\).
}
- The TAC for the management area for 2016 was doubled compared to 2015 but was reduced for 2017 and increased again slightly for 2018 and 2019.
- The decision to exclude discards in the assessment is based on the uncertainty in the available discards data and unknown discard survival rate of plaice.
- A landing obligation is being phased in between 2019 and 2021 for plaice in 7.e with a discard plan. Prior to the introduction of the landing obligation, a substantial part of plaice in 7.e catches has been discarded and not accounted for in the stock assessment.

\section*{8. General Comments}
- The WG report is well written, organized, and follows the annex. The data and data issues are well documented and discussed in detail. The WG did an excellent job providing the sources of uncertainty and bias in the assessment.
- The WG did extensive work for this year's assessment including running an exploratory assessment with a full catch time series and discards for the first time.
- The assessment results depend on assumptions on the mixing rate, which is estimated from a 2010 tagging survey. The RG agrees that spawning structure and mixing rate between 7.d and 7.e need to be investigated and updated to determine if the current mixing rates are still valid assumptions given a general increase in plaice stocks in the English Channel in recent years.
- The RG suggests running a sensitivity analysis with discard data.
- The WG identified an issue related to biological parameters, where natural mortality is borrowed from values of plaice in 7.a stocks but these values have changed since the last benchmark of 7.a. The RG agrees that the biological parameters need to be updated.
- The RG appreciates the recommendations for the next benchmark including the development of a discard time-series into the assessment as discarding was substantial in recent years.

\section*{9. Technical Comments:}
- Figures 23 and 22 are separate and need separate captions.
- Figure 23.24 should read "discard estimate" in legend.
- Short-term projections: "The fishing mortality derived from XSA is above FmSY and the precautionary buffer has never been applied since this stock is treated as category 3, therefore the precautionary buffer should be applied and reduce the advised catch by \(20 \%\)." It is not clear why the precautionary buffer should be applied from this statement.
- Figure 23.4 figure title is missing.
- The RG would like Figures 23.11 and 23.12 to be easier to read.
- The scientific name is not listed under stock description and is only mentioned in the TAC table.
- The RG would like clarification on the FSP survey (definition and description).
- The RG would like clarification about how the tagging data were incorporated into the model.

\section*{10. Conclusions:}
- The assessment of plaice in Western English Channel (ICES division 7.e) appears well done based on all the available information. The RG appreciates the WG's detailed recommendations for the next benchmark. The RG accepts this assessment.
ple.27.7fg. Plaice (Pleuronectes platessa) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea)
1. Assessment Type: Update
- Benchmark: 2011
2. Assessment: Accept with caveats
3. Forecast:
- Short term projection - SPiCT
- No medium or long term was carried out.
4. Assessment Method: Biomass dynamic model (SPiCT-Stochastic Production model in Continuous Time).
5. Consistency:
- The model configuration is different from the 2011 benchmark, though the new model fits the stock better.
- This model is the same as that used in the 2018 assessment.
6. Stock Status:
- SSB is estimated to have been increasing between 2005 and 2018 and began to decline in 2018.
- The 2018 catch was \(421.7 \mathrm{t}, \sim 17.5 \%\) below the TAC of 511 t .
7. Management and Biological Reference Points:
- No management plan has currently been established for Celtic Sea plaice.
- TAC for 2019 is 1662 t where the TAC for 2018 was 511 t .
- There was no early closure of the fishery in 2018.
- The uncertainty cap was not applied.
- Current BRPs:
- \(\mathrm{B} / \mathrm{B}_{\mathrm{MSY}}=1.586\)
- \(\mathrm{F} / \mathrm{F}_{\text {MSYS }}=0.276\)

\section*{8. General Comments:}
- The WG report is well written and follows the stock annex.
- The RG would like to know if there are environmental varibles influencing the stocks.
- Much of the report may need to be updated slightly to newer years and data. These include the input data, survey data, and the assessment diagnostics.
- More solid reference point interpretations and rational in the report may be helpful. Currently most presented information from the report is not discussed in depth.
- The RG suggests the WG discuss the robustness of the currently used model with more detailed interpretation of diagnostics.

\section*{9. Technical Comments:}
- Units are needed on the TAC tables.
- The assessment table shows data being used up to 2017, but it is not clear if more recent data were used in the assessment.
- Many figures need edits:
- Figure captions or descriptions are needed.
- Figure 28.4 title needs edits
- Figure 28.7 is small and hard to read, and the legend does not match the figures.
- Figure 28.10 is small and hard to read, and the legend does not match the figures.
- Figure 28.11 legend does not match the figure.
- Figures 28.10 and 28.11 are not clear.
- Figures 28.8 and 28.9 are missing.

\section*{10. Conclusions:}
- The assessment of Plaice (Pleuronectes platessa) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea) seems to follow the annex well. The RG suggests this report accepted with caveats.
- The environmental vulnerability of this species remains unclear.
- The RG would like to suggest the WG provide more detailed introduction to the input data and interpretations of assessment diagnostics. Many figures and tables attached to the report were not discussed in depth.

Plaice (Pleuronectes platessa) in divisions 7h-k (Celtic Sea South, southwest of Ireland)
1. Assessment Type: Update
2. Assessment: Accept with caveats
3. Forecast: No forecast was carried out.
4. Assessment Method: XSA (Category: 3.2.0)

\section*{5. Consistency:}
- Estimation methodology for MSY reference points changed in 2017, from a data-limited approach to WKMSYREF4.
6. Stock Status:
- SSB increased from \(60 t\) in 2018 to 109 t in 2019 but was below \(\mathrm{B}_{\lim }\) and \(\mathrm{B}_{\mathrm{pa}}\). F bar was 0.68 in 2018 and was above \(\mathrm{F}_{\mathrm{MSY}}\) and \(\mathrm{F}_{\mathrm{pa}}\).
- Total landings of plaice in divisions 7h-k have been declining with fluctuations since 1993 and were 95 t in 2018.
- Discarding is significant but has not been quantified.
- Landings of young fish show a decreasing trend, but it is unclear whether this decrease is due to increased discarding or poor recruitment.
- Recruitment dramatically declined since 1993 but increased from 26,000 in 2018 to 216,000 in 2019.

\section*{7. Management and Biological Reference Points:}
- Plaice in divisions 7h-k are managed by the TAC system.
- TAC set for 2019:
\begin{tabular}{|c|c|c|c|}
\hline Species: \begin{tabular}{l} 
Plaice \\
Plauronetes platessa
\end{tabular} & & Zone: & \(7 \mathrm{~h}, 7 \mathrm{j}\) and 7 k (PLE/7HJK.) \\
\hline Belgium & 7 (1) & & \\
\hline France & 14 ( \({ }^{1}\) ) & & \\
\hline Ireland & 47 (1) & & \\
\hline The Netherlands & 27 (1) & & \\
\hline United Kingdom & 14 ( \({ }^{1}\) ) & & \\
\hline Union & \(109{ }^{(1)}\) & & \\
\hline TAC & \(109{ }^{(1)}\) & & \begin{tabular}{l}
Precautionary TAC \\
Article 8 of this Regulation applies Article 13(1) of this Regulation applies
\end{tabular} \\
\hline
\end{tabular}
\(\left.{ }^{( }{ }^{( }\right)\)Exclusively for by-catches of plaice in fisheries for other species. No directed fisheries for plaice are permitted under this quota.
- ICES advises that when the precautionary approach is applied, there should be zero catches in 2019.
- BRPs were derived in 2017:
\begin{tabular}{|c|c|c|c|}
\hline Framework & Reference POINT & Value & Technical basis \\
\hline \multirow[t]{2}{*}{MSY approach} & MSY B \({ }_{\text {trigger }}\) & 282 & \multirow[t]{7}{*}{\begin{tabular}{l}
\(\mathrm{B}_{\mathrm{pa}}\) \\
Median point estimates of Eqsim with segmented regression \(S-R\) relationship Break point segmented regression \(S-R\) relationship \\
\(\mathrm{B}_{\lim } \mathrm{x} \operatorname{esp}(1.645 \mathrm{x} \sigma) ; \sigma=0.20\) \\
F with \(50 \%\) probability of \(\mathrm{SSB}<\mathrm{B}_{\text {lim }}\) \\
\(\mathrm{F}_{\lim } \mathrm{x} \exp (-1.645 \mathrm{x} \sigma) ; \sigma=0.20\)
\end{tabular}} \\
\hline & FMSY & 0.289 & \\
\hline \multirow[t]{4}{*}{Precautionary approach} & \(\mathrm{B}_{\text {lim }}\) & 203 & \\
\hline & \(\mathrm{B}_{\mathrm{pa}}\) & 282 & \\
\hline & Flim & 0.471 & \\
\hline & \(\mathrm{F}_{\mathrm{pa}}\) & 0.339 & \\
\hline \multirow[t]{2}{*}{Management plan} & \(\mathrm{SSB}_{\text {mgt }}\) & Not applicable & \\
\hline & Fmgt & Not applicable & \\
\hline
\end{tabular}

\section*{8. General Comments:}
- The WG report is well written overall. The data and uncertainty issues are discussed well. Recommendations made by the WG on the future benchmark and management can make an impact on the stock. However, the RG noticed that most of the report seems to be the report from 2018 and the assessment results are presented without a detailed explanation of stock status in the text. The RG encourages the WG to update stock status based on assessment results.
- Although the retrospective analysis was conducted and presented in the assessment report, the methodology was not explicitly mentioned in either the annex or the assessment report. The RG suggests the methodology of retrospective analysis should be at least briefly mentioned in the annex, including the justifications of number of peeling years, specific formulation of Mohn's Rho, and criteria for if the relative quantities (e.g., SSB) should be adjusted based on the outcome of retrospective analysis. Both Mohn's Rho values of SSB and recruitment appear higher than the threshold of \(20 \%\) imposed by ICES for 2019 assessment. However, the WG did not discuss how to deal with the bias.
- As the discard of the stock is significant and it is not included in the present assessment, the RG suggests that more exploratory assessment runs should be performed to investigate the uncertainty (e.g., sensitivity analysis with different discarding scenarios given the currently available data). It is also worthwhile to explore an alternative model that is less sensitive to catch data, such as ASAP, in the future assessment.
- Both the management plan and HCR for the stock are not clear. The RG recommends this section have a more detailed description (not just the tables).
- The stock annex has not been updated since 2014 and is not consistent with the current report. The RG recommends updating of the annex, even in the absence of a benchmark. More details of survey designs for the commercial tuning index should also be included in the annex.
- The RG appreciates the WG's extensive efforts in checking the sensitivity of the tuning index with respect to spatial stratification of fishing efforts. As the WG mentioned, the precision of stratified estimates was relatively low and needed
improvement in the future benchmark. The RG agrees that more sophisticated modelling approach and structured sensitivity analysis should be performed in the future.
- As the stock-recruitment relationship is not clearly defined, recruitment may be strongly influenced by environmental factors and stock mixing. The RG suggests that it is worthwhile to explore the relationship between recruitment and the environmental factors, as well as the interaction with plaice in division 7e.
- Natural mortality is assumed to be fixed and 0.12 for all ages. The RG suggests adding more biological clarifications to justify the choice of the fixed natural mortality, coupled with a structured sensitivity analysis to evaluate impacts of uncertainty in natural mortality.
- The RG agrees with the WG's suggestion regarding the use of new datasets from Irish surveys as an updated tuning index. However, the RG suggests the consistency of survey designs should be carefully considered in the future assessment. The current assessment is primarily on the stock in divisions 7 j and 7 k , which is unrepresentative of the whole stock area (7h-k). The RG agrees to use age-structured data in 7h for the future benchmark.

\section*{9. Technical Comments:}
- The pdf link for ICES advice applicable to 2019 is unavailable and needs to be updated.
- "This distance would suggest that \(7 . h\) stock may constitute a spate stock, and may be a continuation of the plaice caught in the western English Channel (7.e)." The "spate stock" seems to be a typo and needs to be corrected.
- "With plaice forming only a small component ( \(<5 \%\) ) of the overall landings per trip (Figure 27.1)" Figure 27.1 does not support the statement and needs to be updated.
- The WG should provide the unit for both TAC tables for 2018 and 2019.
- "Irish Beam Trawl Ecosystem Survey (IBES)" There are two forms of abbreviation: IBES and IBTS. Please make sure they are consistent throughout the report.
- "A summary of relative trends in landings, recruitment, SSB and F is given in Table 27.10 and Figure 27.7". It should be "recruits" instead of "recruitment". Table 27.10 is missing. Figure 27.7 does not provide any trend information.
- "Details on this evaluation can be found in the working document in appendix xxx" Please specify the appendix.
- "From this Blim was estimated to be 203.57 (Blim <- median(fit\$sr.sto\$b.b)) and a Bpa at 282.88 (Bpa <-Bpa(Blim, 0.2)." Please provide the unit of Blim and Bpa.
- "Figures 7.11.12 and 7.11.13 summarise the MSY evaluation.". These two figures are missing.
- "The apparent reduction in SSB in 2015 is mainly driven by a reduction in relative abundance of young fish in recent years, there is a slight increase in 2016, but is
again showing a downward trend in 2018 " No evidence is provided with this statement. The RG encourages the WG to present supporting data and materials to justify the statement.
- "It is unclear, whether this lack of young fish in the landings (and commercial tun-ing lpue index) is due to increased discarding or poor recruitment (Table 27.1)." Table 27.1 does not present the landings of young fish. Please update it and ensure consistency between the text and the table.
- "Because plaice are caught in spatially distinct areas, restricting effort in these areas will be more effective than limiting landings." The RG is not clear how the effort restriction is more effective than landings regulations. Please justify this statement.
- "It is likely that the plaice from Division 7.h are part of the divisions 7.e or 7.fg stocks." Please provide supporting references.
- Tables and Figures need to be listed in descending order according to their number.
- Table 27.5 - The ages and years used in the assessment are not marked in bold.
- Table 27.7 - It should be "recruit" instead of "recruitment" in the table caption. Blanks cells need to be filled with NAs.
- Figure 27.1.a and Figure 27.1.b should be Figure 27.1 and Figure 27.2 respectively. Figure 27.1.b needs to be updated with landings in 2018 and 2019. Please consider adding a figure of landing trends in divisions 7.h and 7.k.
- Figure 27.4 - The size of bubbles needs to be clarified in the figure caption.
- Figure 27.7- The legend is not complete.
- Figure 27.11 - The WG needs to provide the unit of \(y\)-axis. A legend of lines is also missing.
- Figure 27.12 and Figure 27.13 - X-axis needs to be labeled. The unit of y-axis is not provided. The caption is not clear, e.g., does \(b\) represent SSB or biomass?

\section*{10. Conclusions:}
- In general, the WG report is well written. Data and data issues are well documented and stated. However, the retrospective analysis shows large systematic bias and was not explicitly discussed by the WG. The RG suggests adding a correction. More exploratory assessment runs need to be conducted to investigate uncertainties, especially in discards, natural mortality, and stock-recruitment relationship.
- The current assessment is primarily on the stock in divisions 7 j and 7 k , which is considered unrepresentative of the whole stock area ( \(7 \mathrm{~h}-\mathrm{k}\) ). As age-structured data became available for 7 h , the RG recommends adding quantified analysis to illustrate the uncertainty.
- A detailed explanation of stock status and management plan is lacking. The RG suggests including more discussion of stock status based on the current assessment and clarifying management plan for the stock.
- The RG suggests assessment of plaice in divisions 7h-k to be accepted as long as the above concerns are addressed.
pol.27.67. Pollack (Pollachius pollachius) in subareas 6-7 (Celtic Seas and the English Channel or West of Scotland)
1. Assessment Type: Update (category 4 stock, no benchmark)
2. Assessment: Accept with caveats
3. Forecast: N/A
4. Assessment Method:
- Depletion-Corrected Average Catch (DCAC) method was used.

\section*{5. Consistency:}
- The method was applied during WGCSE 2019 with the same model settings as applied the previous year's assessment
- Subarea 6 and 7 are run independently.
- Discarding is negligible.
- Recreational catch is unknown and therefore cannot be estimated or incorporated into the assessment. In 2018, the recreational catches may be similar to or above commercial landings.
- A model test is constructed in the WG report which highlights that the DCAC model will not take any account of the state of the stock.
- The results are consistent with the range of DCAC values estimated when the method was previously applied.
6. Stock Status:
- Stock status is unknown due to the lack of reliable data.
- Commercial catches have declined since the late 1980's, and in 2018 is the lowest in the time-series.

\section*{7. Management and Biological Reference Points:}
- The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters. This plan applies to demersal stocks including Pollack in ICES subareas 6 and 7.
- ICES advices that when the precautionary approach is applied, commercial catches should not exceed 3360 tonnes in 2020.
- The WG suggests that yield in subarea 6 could be increased up to 148 tonnes and 4010 tonnes in subarea 7.

\section*{8. General Comments:}
- The WG report is well written. The data and data issues are well documented and discussed at length.
- The RG suggests that a stock annex for pol. 27.67 should be developed.
- The stock in the WG report relates to a species in a water region where data are available. The stock structure of Pollack populations in this eco-region is not clear. The RG suggests that further work is required for identifying a management unit.
- The use of DCAC is not recommended if M is greater than 0.2 , above which the depletion correction becomes small. Natural mortality is fixed at \(\mathrm{M}=0.2\) with a standard deviation of 0.4 in the WG report. The RG suggests adding more biological clarification to justify the choice of natural mortality.
- The result from the model tests highlights that the DCAC model will not take any account of the state of the stock. The RG suggests that the purpose, assumption, and the process for testing the uncertainties should be shown more in detail.

\section*{9. Technical Comments:}
- There is no scientific term for the species and the format of the species name is not uniform in the report. The RG suggests providing the scientific term and unifying the format of species name.
- No caption for the first figure in the report. The RG suggests providing the figure title or change the figure into a table with a title.
- Section 28.1-Fishery in 2018- Landings by division-"In subarea 7, the division with the highest proportion of landings derived from 7.e (40\%) followed by 7.j (27\%), 7.h (13\%) and 7.f (9\%). Landings in divisions 7.a, b, c, d, g and h were negligible ( \(8.9 \%\) )". There are two 7.h. One of them may be 7.k. The RG suggests to make sure which 7.k is.
- The RG suggests discarding from subarea 6 and 7 should be estimated independently. The total estimated discarding is at 19.9 tonnes in 2018. It might not be a small part of the landings for subarea 6.
- The RG suggests using the three-line forms for the table in WG report.
- The RG suggests providing the meanings of the parameters in table 28.3 and 28.4 and explain the basis for setting these parameters
- The text use "tonnes", as a unit label, while in figure 28.1 uses "1000 tons". Keep the unit label consistent and use "tons" as the unit.
- In Table 28.1, 28.2, 28.3, 28.4, Unit label of the landings should be provided.

\section*{10. Conclusions:}
- In general, the WG report is well written and organized. The data and data issues are well documented and discussed at length.
- The results are consistent with the range of DCAC values estimated when the method was previously applied. The RG recommends this stock assessment to be accepted.
- As the WG noted, the model cannot estimate reference points and it continues to be the reference model for the pollack assessment in the coming year. The RG suggests that the recreational fishery should be monitored and agrees that future assessments should consider exploring new models.

\section*{Sole (Solea solea) in Division 7.a (Irish Sea)}
1. Assessment Type: Update
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- Short term projection - Carried out by an age structured deterministic projection. The geometric mean is assumed for age 2 in the short-term forecast. The procedure for setting the fishing mortality is to take the mean over the last three years, without rescaling.
- Medium-term projection - Carried out by an age structured model, using IFAP single option prediction software.
- Long-term projection - Carried out by an age structured deterministic projection, using MFYPR software. The inputs for the long-term projections are the same as for the shortterm projection.
4. Assessment Model: Extended Survivor Analysis (XSA) was used.

\section*{5. Consistency:}
- The model configuration is the same as the 2018 assessment, but since the last benchmark (2010) these have been some changes, as follows: the time interval of the UK Sept BTS was changed to reflect the actual period for the 2019 survey, the UK Mar BTS was omitted in the 2019 survey, the time ser. Wts in the 2019 benchmark have been changed to no taper weighting, and the Q plateau has been changed from 7 to 4 .
- Fishing mortality, SSB and recruitment displayed trends that were very similar. In last year's assessment, F and SSB for 2017 were estimated to be 0.0188 and 1941 t respectively; this year's estimates for 2017 are 0.0193 and 1891 tonnes.
- This year's model output suggested a downward revision of the historical trends for SSB, and an upward revision of F and recruitment.
- A TAC constraint was recommended for 2019 as SSB rose above \(\mathrm{B}_{\mathrm{lim}}\). As the stock is projected to slowly recover but remain close to \(\mathrm{B}_{\mathrm{lim}}\) in 2020, there is a further increase in non-zero catch advice for the year 2020.

\section*{6. Stock Status:}
- The current SSB is below \(\mathrm{B}_{\mathrm{pa}}\) that is set at 3500 tonnes. SSB has increased from 883 tonnes in 2014 to 3079 tonnes in 2019. The SSB is predicted to be 3204 tonnes in 2020, when \(\mathrm{F}=0.138\).
- Since the 1980s, F has been declining and in 2018, F was close to 0 , at 0.0138 . F For \(2020=0.29\).
- Total catch in 2018 was 38 tonnes while the TAC was 40 tonnes.
- Catches have steadily decreased since the late-1980s.
- SSB peaked in the mid-1980s and was well below \(\mathrm{B}_{\text {lim }}\) between 2004 and 2017.
- Since 1970, recruitment was at an all-time low during 2011-2014. Since then, recruitment has increased above recent average at 3670 tonnes in 2018.

\section*{7. Management and Biological Reference Points}
- Biological reference points were re-evaluated in 2015.
- \(B_{l i m}\) and \(B_{p a}\) have increased from \(2200 t\) to \(2500 t\) and \(3100 t\) to \(3500 t\) respectively. \(B_{p a}\) was decided by using the lowest value with above-average recruitment. \(\mathrm{B}_{\text {lim }}\) was calculated by taking the approximate \(\mathrm{B}_{\mathrm{pa}}\) value and multiplying it by 1.4.
- \(\mathrm{F}_{\mathrm{pa}}\) and \(\mathrm{F}_{\text {lim }}\) have decreased from 0.3 to 0.21 and from 0.4 to 0.29 respectively.
- Catches in 2020 should be no more than 558 tonnes.
- No minimum landings size has been defined.
- MSY advice is currently based upon stochastic simulations with a segmented regression stock-recruitment relationship, while \(\mathrm{B}_{\text {trigger }}\) is considered as equivalent to \(\mathrm{B}_{\mathrm{pa}}\).
- Current biological reference points;
- \(\mathrm{B}_{\mathrm{lim}}=2500\) tonnes
- \(\mathrm{B}_{\mathrm{pa}}=3500\) tonnes
- \(\mathrm{F}_{\text {lim }}=0.29\)
- \(\mathrm{F}_{\mathrm{pa}}=0.21\)
- \(\mathrm{F}_{\max }=\mathrm{NA}\)
- \(\mathrm{F}_{0.1}=\mathrm{NA}\)
- \(\mathrm{F}_{\text {MSY }}=0.20\)
- \(\mathrm{F}_{\text {MSY }}\) lower \(=0.16\)
- FMSY upper \(=0.24\)
- MSY \(\mathrm{B}_{\text {trigger }}=3500\) tonnes

\section*{8. General Comments:}
- The report is well written and follows the stock annex. The data and data issues are documented clearly and discussed at length. The tables provided were labeled clearly and were easy to read. The WG also did a great job on providing the necessary information while also keeping the report concise.
- The RG supports the WG in their recommendation to study how environmental changes are impacting the distribution of sole. The RG suggests that the WG incorporate environmental variables into the estimation of the survey abundance index.
- The RG suggests that the WG conduct a sensitivity analysis on the discards data and assuming different discard rates so that the WG can compare and interpret the results of the comparisons.
- A limitation with using this assessment is that the prediction ability may not always the best (Sole-bisc review report from 2018). The RG suggests that the WG try different
settings or parameters and to cross-validate to find the scenario that produces the best prediction output of the model.
- The WG noted that different effort unit results were used in different trends of LPUE. The RG is wondering if both units were tried, and what the comparison of the results would look like.
- The WG could provide more insight into the sources of uncertainty relating to this stock, such as how natural mortality was assumed to be constant even though there are no estimates of natural mortality to provide data, as mentioned in section B. 2 of the stock annex. Another source of uncertainty that could be elaborated on are the low sampling level and sample size of sole in recent years and how that could be altering the results of the assessment. The WG could conduct a few more exploratory runs to address the issues regarding any uncertainties. On that note, the WG mentioned that in recent years, both sampling level and sampling size have been low. The RG suggests that the WG conduct a simulation study to test whether the low sample sizes would accurately reflect the size distribution and age structure of sole, as imprecise data could affect the assessment results. In addition, the WG could conduct a sensitivity analysis considering these uncertainties of input data. The WG also noted that there were particularly low sampling levels in the first quarter. The RG suggests that the WG consider using different weights from different quarters.
- Though a five-year retrospective analysis was performed, which concluded the assessment to have a high consistency, the RG suggests conducting a longer retrospective analysis to better cover the lifespan of sole.

\section*{9. Technical Comments:}
- The WG should write out "Extended Survivor Analysis" before writing "(XSA)", at least for the first time the model type is mentioned.
- The description of Figure 1 does not mention the Catches graph.
- It would help clarify if the \(\mathrm{F}_{\text {lim }}, \mathrm{F}_{\mathrm{pa}}\), and \(\mathrm{F}_{\text {MSY }}\) reference points on the Fishing Mortality graph, as well as the \(\mathrm{B}_{\mathrm{lim}}, \mathrm{B}_{\mathrm{pa}}\), and MSY \(\mathrm{B}_{\text {trigger }}\) reference points on the SSB graph in Figure 1 had their corresponding reference point numbers next to the labels in the legends of the graphs. Also make sure the appropriate letters are subscripted on the legend I Figure 1. For example: "FMSY" instead of "FMSY".
- Section B-2-"Males and females of this stock are strongly dimorphic, with males showing much reduced rates of growth after reaching maturity, whilst females continue to grow. Given the minimum landing size of 24 cm the majority of landings represent mature females." The RG suggest that the WG provide references to support this statement.
- The RG suggests making Table 1 larger, if possible so that it is easier to read.
- Figure 2 needs axis labels on graphs. The SSB and Fishing mortality graphs in Figure 2 need legends that provide labels for \(\mathrm{F}_{\mathrm{lim}}, \mathrm{B}_{\mathrm{lim}}, \mathrm{F}_{\mathrm{pa}}, \mathrm{B}_{\mathrm{pa}}, \mathrm{F}_{\mathrm{MSY}}\), and MSY \(\mathrm{B}_{\text {trigger }}\).
- "The assessment has shown consistency over the recent years in estimating SSB, fishing mortality, and recruitment. Discards are currently not included in the assessment, but given the low discard rates of sole (3.5\% in 2016-2018) it is unlikely that the inclusion of discards would change the perception of the stock." No comma is needed between "assessment" \& "but".
- "The stock is slowly recovering and is projected to remain close to Blim in 2020 and 2021, and may fall below \(B_{\text {lim }}\) if recruitment is below average" No comma is needed between "2021" \& "and".
- At the bottom of table 7, some cells were left blank. It would be helpful to either mark them as \(0, N / A\), No data, or with any other appropriate entry, as blank cells could become confusing to the audience. It is recommended no cells be left blank in any table.
- The RG found the note at the bottom of table 10 " \(\operatorname{GM}(2009-2017)\) " to be unclear. The RG suggests that the WG define what "GM" stands for.
- In the stock annex, it was stated "For 2009 Council Regulation (EC) Nº43/2009 allocates different amounts of \(k W^{*}\) days by Member State and area to different effort groups of vessels depending on gear and mesh size. The areas are Kattegat, part of IIIa not covered by Skaggerak and Kattegat, ICES zone IV..." The RG suggests the WG provide a map of the coverage of Kattegat, Skaggerak, ICES zone IV, etc.
- Figure 29.4a: It would be helpful to describe the effort that was used to represent the graph, perhaps in a description of the figure.
- Figure 29.6: Explain legend on the right (2-7 represent age?). Also, "Standardised" in figure title is spelt incorrectly.
- Figure 29.9: Need y axis labels, no underscores in titles of graphs.
- Figure 29.10: Needs axis labels, could add separate x axis to each graph to make clearer.
- Figure 29.11: Graphs need axis labels with units on all.
- Figure 29.12: Percentages do not sum to 100.
- Table 29.3: "Length" is misspelled in title.

\section*{10. Conclusions:}
- The assessment of Sole in division 7.a appears well done and indicates no large retrospective errors. Sources of uncertainty could be discussed in greater detail. Minor formatting and grammar errors should be revised. The RG suggests sole in division 7.a to be accepted if the following concerns are addressed: a) different parameter settings are tried and cross validated to ensure that the best prediction scenario of the model is being used; b) a simulation study be conducted to test whether the low sample sizes would accurately reflect the size distribution and age structure of sole.

\section*{Sole (Solea solea) in Division 7.e (Western English Channel)}
1. Assessment Type: Update
2. Assessment: Accepted

\section*{3. Forecast:}
- Short-term projection - performed in the XSA forecast based on MSY approach for 2019-2021.The SSB was estimated to be 4756 tonnes in 2019, 4731 tonnes in 2020, and 4334 tonnes in 2021.The landings was estimated to be 1216 tonnes in 2019, 1469 tonnes in 2020, and 1349 tonnes in 2021.
- Mid-term projection: N/A
- Long-term projection: N/A
4. Assessment Method: Extended Survivors Analysis (XSA)

\section*{5. Consistency:}
- The last benchmark assessment was in 2012 and the assessment was interbenchmarked in 2015 for examining the impact of excluding UK Wester Channel beam trawl survey data and revising the model setting to increase the robustness of the assessment.
- Seasonal onshore-offshore spawning migrations were briefly mentioned in the stock annex; however, spawning season as well as fishery seasonality were not clear in the stock annex. The RG expects to see more information regarding life history of sole in 7.e and their relation to fishery.
- There appeared to be changes in fishing patterns in recent years where the fishery opted for smaller and more flexible vessels which allow fishermen to exploit other species in this Division. The RG suggests the changes in fishing patterns in recent years and the impact of that on sole stock should be evaluated.
- A new management plan, EU multiannual plan (MAP), is implemented in 2019. And the landing obligation also fully applies in 2019. The RG suggests the stock annex should be updated regarding the recent changes in fishery management.
- The maturity ogive used for this sole stock was a maturity ogive from area 7.f and 7.g estimated in 1997. The RG would like to see more information on maturity as different stocks may present different life history traits. It might not be appropriate to use the maturity ogive from other stocks estimated two decades ago, although it is not an urgent concern as the sole stock in this division seems to have been in healthy status.
- The methodology of how effort, lpue, and cpue were standardized seemed to be missing from the stock annex and the assessment report. The RG suggests it should be added to the documents.

\section*{6. Stock Status:}
- MSY \(_{\text {trigger }}=B_{p a}=\) MAP \(B_{\text {trigger }}=\) MAP \(B_{p a}=2900\) tonnes.
- \(\mathrm{B}_{\mathrm{lim}}=\) MAP \(\mathrm{B}_{\mathrm{lim}}=2000\) tonnes
- \(\mathrm{F}_{\mathrm{MSY}}=0.29 ; \mathrm{F}_{\mathrm{pa}}=0.32 ; \mathrm{F}_{\text {lim }}=0.44\)
- MAP range \(\mathrm{F}_{\text {lower }}=0.16, \mathrm{MAP}\) range \(\mathrm{F}_{\text {upper }}=0.34\)
- SSB has been increasing since 2009 and is currently well above MSY \(\mathrm{B}_{\text {trigger }}\)
- Recruitment variability has decreased since 1991, and has been at or above the long term geometric mean in the last four years.
- Fishing mortality has remained low since 2009 and stayed at around \(0.2-0.25\). Fishing mortality was estimated to be well below all reference points.

\section*{7. Management and Biological Reference Points:}
- The EU landing obligation which aims to reduce unwanted catches fully applies to sole in Division 7.e in 2019. However, given the low discards observed in the fishery, this management plan is not expected to have a considerable impact on this sole stock.
- A new management plan (EU multiannual plan) is implemented in 2019. This management plan imposes the effort restrictions on the number of days at sea for certain vessels.

\section*{8. General Comments:}
- The WG report is well written and followed by the Terms of References. The input data sources were clearly provided and most data issues were also discussed in the stock annex and the assessment report.
- The assessment model was tuned with two survey indices (UK-FSP and Q1SWBeam) and lpue time series from two commercial fleets (UK-CBT and UK-COT). However, the UK-COT effort had been in continual decline since 1970s and zero effort and landings were reported since 2016, a result of the shift to smaller fishing vessels. Although a new data base has been used since 2017, the new data were not consistent with historical data. The RG suggests the effects of changing commercial fishing vessel sizes should be further examined and taken into account into the assessment. Furthermore, the impact of the loss of information derived from UK-COT which seemed to provide a good independent time series from the main commercial catches should also be evaluated. The inconsistency for new data base also needs to be examined.
- The impact of fishing effort restriction implemented by the new EU multiannual plan should be evaluated and the consequent changes in fishing pattern should also be examined.
- The RG would like to see further discussion on residuals diagnosis regarding model fitting.
- The RG agrees that assessment parameterization may need to be re-evaluated in the future when the tuning time series get longer.

\section*{9. Technical Comments}
- Table 31.6 has some numbers that seem to be skewed in relation to the table and where they are supposed to be. Perhaps a revision of the table format is necessary.
- Figure 31.20.What is WG year?
- Figure 31.21.What is fshk? Was it mentioned in the assessment report? The tick labels on \(x\)-axis were not clear.
- Figure 31.22. The RG suggests the time series in the figures should be standardized to help visualize the deviation from the full model.
- Figure 31.24. What is STF?
- Figure A2.1 in stock annex. The x-axis is not clear.

\section*{10. Conclusions}
- The assessment for sole in Division 7.e appears well done with the revised XSA model.
- The major issues of the input data had been discussed in the stock annex, however the impact of the loss of information from UK-COT may need to be further evaluated.
- The WG may need to examine the impact of changes in fishing pattern on sole and the consequences resulted from the new management plan, and take these into account for the assessment.

\section*{Sole in divisions 7.f and 7.g (Bristol Channel, Celtic Sea)}
1. Assessment type: updated; next benchmark 2020
2. Assessment: Accepted with caveats

\section*{3. Forecast:}
- Short-term: prediction of SSB in 2020~2021 given certain catches in 2019 and 2020

○ (2019catch: 841t, 2020catch: 1236t; 2020SSB: 5366t, 2021SSB: 5459t)
- Medium-term: no
- Long-term: no

\section*{4. Assessment Method:}
- XSA (Extended Survivors Analysis)
- Main model configurations:
- SSB-R relationship: Segmented regression
- b.2. Natural mortality: the natural mortality was assumed to be 0.1 for sole of all the ages in all the assessment years.
- b.3. The maturity ogive is based on samples taken during the UK(E\&W) beam trawl survey of March 1993 and 1994
- Input data sources:
- c.1. Survey data from UK
- c. 2 Commercial fishery data from UK and Belgium

\section*{5. Consistency:}
- The Mohn's rho values (SSB:0.071; F:-0.05; Recruitment:0.105) for the assessment model are low and well below the threshold of \(20 \%\).
- The internal consistency plots for the commercial LPUE series show high consistencies for the entire age range.
- However, comparing with the last year's stock assessment results, there is a substantial increase in the SSB and a decrease in the F. This may be due to the change of the input data.

\section*{6. Stock Status:}
- In 2018, SSB and F were 3557 tons and 0.229 , respectively. F was below Fpa and SSB was above Bpa.
- Recruitment has fluctuated around 5 million recruits with occasional strong year classes.
- SSB has declined almost continuously from the highest value of 7385 t in 1971 to the lowest observed in the time-series in 1998 (1592 t).
7. Management and Biological Reference Points:
- Reference points
\begin{tabular}{|c|c|c|c|}
\hline Framework & Reference point & Value & Technical basis \\
\hline \multirow[b]{2}{*}{MSY approach} & MSY \(\mathrm{B}_{\text {trigger }}\) & 2228 tonnes & \(\mathrm{B}_{\mathrm{pa}}\) \\
\hline & \(\mathrm{F}_{\text {MSY }}\) & 0.297 & EQsim analysis based on the recruitment period 1971-2017 \\
\hline \multirow{4}{*}{\begin{tabular}{l}
Precautionary \\
approach
\end{tabular}} & Blim & 1592 tonnes & \(\mathrm{B}_{\text {loss }}\) estimated in 2018, corresponding to SSB in 1998 \\
\hline & \(\mathrm{B}_{\mathrm{pa}}\) & 2228 tonnes & \(\mathrm{B}_{\text {lim }} \times 1.4\) \\
\hline & \(\mathrm{F}_{\text {lim }}\) & 0.587 & EQsim analysis, based on the recruitment period 1971-2017 \\
\hline & \(\mathrm{F}_{\mathrm{pa}}\) & 0.420 & \(\mathrm{F}_{\text {lim }} \times \exp (-1.645 \times 0.2) \approx \mathrm{F}_{\text {lim }} / 1.4\) \\
\hline \multirow{6}{*}{Management plan*} & MAP
MSY B \(_{\text {trigger }}\) & 2228 tonnes & MSY \(\mathrm{B}_{\text {trigger }}\) \\
\hline & MAP \(\mathrm{B}_{\mathrm{pa}}\) & 2228 tonnes & \(\mathrm{B}_{\mathrm{pa}}\) \\
\hline & MAP \(\mathrm{B}_{\text {lim }}\) & 1592tonnes & \(\mathrm{B}_{\text {lim }}\) \\
\hline & MAP F MSY & 0.297 & \(\mathrm{F}_{\text {MSY }}\) \\
\hline & MAP range \(\mathrm{F}_{\text {lower }}\) & 0.165-0.297 & Consistent with ranges provided by ICES (2019a), resulting in no more than \(5 \%\) reduction in long-term yield compared with MSY \\
\hline & MAP range \(\mathrm{F}_{\text {upper }}\) & 0.379-0.499 & Consistent with ranges provided by ICES (2019a), resulting in no more than \(5 \%\) reduction in long-term yield compared with MSY \\
\hline
\end{tabular}
- Celtic Sea sole has been managed by TAC since 1983
- Technical measures including minimum landing size and minimum mesh sizes
- Temporal closures

\section*{8. General Comments:}
- The average discard rate for 2016-2018 is \(8.9 \%\), while the discard rate for 2018 is \(14.8 \%\). We recommend considering discard in the stock assessment model as it becomes a relatively important component of the total catch.
- The natural mortality was assumed to be 0.1 for sole of all the ages in all the years. This strong assumption needs more justification or the consequences of violation of the assumption need to be understood by conducting sensitivity analysis.
- Please justify the reason for not including some available data (e.g., abundance indices of the Irish ground-fish survey) in the model.
- The RG appreciates that the WG tuned the abundance index from the commercial fishery with the factor of engine power. The RG believes that the quality of the abundance index data could be further improved by conducting an LPUE standardization analysis which is aimed to exclude impacts on nominal LPUE other than stock abundance.
- The uncertainty of the estimates of the biomass and recruitments should be quantified.
- The RG agrees that alternative models (e.g., ASAP, SAM) should be explored in the next benchmark stock assessment.
- It will be great to provide the justification of the precautionary approach and the reference points setting as well as the clear description of the harvest control rule.

\section*{9. Technical Comments:}
- "the proportional contributions of recent year classes to the predicted landings and SSB are given in Figure 36.15. The assumed GM recruitment accounts for about \(2.3 \%\) of the landings in 2020 and about \(11.3 \%\) of the 2021 SSB." The description is not consistent with Figure 36.15. Specifically, the percentage \(11.3 \%\) is higher than the data shown in the figure.
- Figure captions are missing

\section*{10. Conclusions:}
- The assessment of Sole in divisions 7.f and 7.g is well organized. In this year's update stock assessment, the stock is assessed using an XSA model, which is one version of virtual population analysis.
- The RG suggested conducting LPUE standardization analysis to improve the reliability of the fishery dependent data. The consequence of ignoring the discard or other observation errors associated with the catch data should be understood.
- The RG agrees that alternative models (e.g., ASAP, SAM) should be explored in the next benchmark stock assessment.
- It will be great to provide the justification of the precautionary approach and the reference points setting as well as the clear description of the harvest control rule.

Sole (Solea solea) in Division 7.h-k (Celtic Sea South, southwest of Ireland)
1. Assessment Type: Update
- Last benchmark: unknown
2. Assessment: Accepted with Caveats

\section*{3. Forecast:}
- Short-term projections: N/A
- Mid-term projections: N/A
- Long term projections: N/A
4. Assessment Method: Extended Survivors Analysis (XSA)
- Follows the Annex
- Commercial landings data
- Survey data starting in 2016
- Annex says survey may not even be representative because of highly variable catchability of sole with gear type.

\section*{5. Consistency:}
- It is unclear when the last benchmark assessment was. It is known that this stock is scheduled to be benchmarked in 2020.
- The information provided in the annex was incomplete and considerable errors were found (e.g. the caption of the first figure stated it was a map for Plaice international landings). The RG suggests the WG to considerably update the annex for the coming benchmark assessment.

\section*{6. Stock Status:}
- SSB declined from 800 tonnes to 400 tonnes 2000-2009, but has recovered to 800 tonnes again in recent years, only to drop to 425 tonnes.
- MSY Btrigger \(=\mathrm{Bpa}=590\) tonnes.
- \(\mathrm{B}_{\text {lim }}=425\) tonnes
- \(\mathrm{F}_{\mathrm{MSY}}=\mathrm{F}_{\mathrm{pa}}=0.161 ; \mathrm{F}_{\text {lim }}=0.222\)
- Current SSB is 460 tonnes.
- Current F is 0.111.

\section*{7. Management and Biological Reference Points:}
- The EU landing obligation. However, given the low discards observed in the fishery, this management plan is not expected to have a considerable impact on this sole stock in these areas.
- A new management plan (EU multiannual plan) is implemented in 2019. This management plan imposes the effort restrictions on the number of days at sea for certain vessels.
- MSY Btrigger \(=\mathrm{Bpa}=590\) tonnes.
- \(\mathrm{B}_{\text {lim }}=425\) tonnes
- \(\mathrm{F}_{\mathrm{MSY}}=\mathrm{F}_{\mathrm{pa}}=0.161 ; \mathrm{F}_{\text {lim }}=0.222\)

\section*{8. General Comments:}
- Considering there is no clear stock-recruitment relationship and given the fact that the recruitment has been declining in recent years and drops to a record low level last year, which could be a sign of stock collapsing, the RG suggests the biological reference points (BRPs) should be determined more conservatively. The RG suggests two alternatives: (1) consider \(\mathrm{B}_{\text {loss }}\) (the lowest SSB where large recruitment was observed) as a \(\mathrm{B}_{\text {lim }}\) candidate. This BRP is usually adopted for stocks which do not have a clear S-R relationship with occasional large year classes observed. (2) consider \(\mathrm{B}_{\text {loss }}\) (the biomass corresponding to the lowest observed SSB ) as a \(\mathrm{B}_{\text {lim }}\) candidate. This BRP is usually adopted for stocks without apparent S-R signal. Either of these two BRPs would be more conservative than the \(\mathrm{B}_{\mathrm{lim}}\) adopted in the current assessment. The RG suggests that \(\mathrm{B}_{\mathrm{lim}}\) should be taken with extra caution as the \(\mathrm{F}_{\mathrm{lim}}, \mathrm{B}_{\mathrm{pa}}\), and \(\mathrm{F}_{\mathrm{pa}}\) are all estimated from \(\mathrm{B}_{\text {lim }}\).
- The WG stated that restricting effort will be more effective than limiting landings as sole in these areas are caught in spatially distinct waters. The RG would like to see more discussion on the expected effects of the new management plan (EU multiannual plan) which imposes the effort restrictions on the number of days at sea for certain vessels for sole stocks in these areas.
- The maturity ogive used for this sole stock seemed to be estimated over two decades ago from other stocks. The RG suggests the maturity ogive needs to be further investigated as different stocks may have different reproductive traits. Furthermore, the declining weight-at-age of older fish (Fig. 33.5) might also have an impact on age-at-maturity.

\section*{9. Technical Comments:}
- Many of the table numbers and Figure numbers were not matched with those in the text.
- What are Irish OTB vessels?
- The WGCSE report section 37.7 has duplicate sentence errors.
- Figure 33.2. It is unclear what are the labels and units on x-axes and \(y\)-axes of these figures. The values in the graph also need to be explained.
- Figure 33.8. Please provide (Pearson's?) correlation coefficients and p-values for each correlation.
- Figure 33.9. Please add legend and scale for black and white circles in this figure.
- Retrospective analysis did not seem to be conducted in this assessment. The output of retrospective analysis in section 37.3.2 appeared to be copied and pasted from other stock assessment report?
- Section 37.4. Please correct the appendix number of document for details on MSY evaluation.
- It is unclear why \(\mathrm{F}_{\mathrm{cv}}\) and \(\mathrm{F}_{\text {phi }}\) used in the assessment were the same as those used for plaice in Division 7.e. Please add explanation.

\section*{10. Conclusions:}
- The RG accepts this report under the following caveats:
- The WG reevaluate all information, tables, and figures present in the report and annex to ensure that the information being presented is true for Sol.7.h-k and not information accidently copied from a different stock.
- The above general and technical comments are addressed, with high attention given to general comments 1 and 2 .

\section*{Whg.27.7a : Whiting (Merlangius merlangus) in Division 7.a (Irish Sea)}
1. Assessment Type: Update (benchmarked in 2017, ICES 2017)
2. Assessment: Accepted with caveats

\section*{3. Forecast:}
- Short term projection - Using FLAssess.
- No medium or long term was carried out.
4. Assessment Method: Age-Structured Assessment Program (ASAP) model was used of basis for advice (V3.0.17 NOAA Fisheries toolbox).

\section*{5. Consistency:}
- The assessment configuration is the same as the 2018. A full analytical assessment procedure was developed during WKIRSH 3 (ICES, 2017) using ASAP.
- There is no new source of data used in this assessment compared to the previous assessment. An additional year of catch and survey data were included in 2018.
- The majority of catches have been discarded for the last couple of decades. Despite increased sampling levels, discard information remains very imprecise. This has contributed to the highly fluctuating fishing pressure estimates in recent years.
- The result of the update assessment is consistent with last year's assessment and indicates that fishing mortality has declined significantly, but the stock size remains extremely low. - This stock is now subject to the landings obligation by way of the Commission Delegated Regulation (EU) 2018/2034 compared to the previous assessment (2018). In 2019, the TAC was increased to 727 tonnes.
- The 2019 assessment includes a retrospective analysis. The Mohn's rho values for this assessment are below the threshold of \(20 \%\) imposed by ICES for 2019 assessments.

\section*{6. Stock Status:}
- Information from commercial data in recent years suggests that the present stock size is extremely low.
- Total catch has increased from 703 tonnes in 2017 to 899 tonnes in 2018.
- Landings have increased from 36 tonnes in 2017 to 46 tonnes in 2018, and reported discard levels appear stable in recent years (Figure 1).
- SSB has been declining since the beginning of the time-series and has been well below \(\mathrm{B}_{\text {lim }}\) since the mid-1990s.
- Recruitment has been low since the early 1990s. Fishing pressure (F) has declined since 2015 but remains above \(\mathrm{F}_{\text {MSY }}\) and \(\mathrm{F}_{\text {lim }}\) in 2018.

\section*{7. Management and Biological Reference Points:}
- No management plan has been agreed or proposed.
- The Whiting fishery is currently managed by TAC and technical measures
- Agreed TAC for 2019 was 727 tonnes ( 2018 was 80 tonnes )
- When the precautionary approach is applied, there should be zero catches in 2020.
- Discarding in the Nephrops fishery is the main management issue.
- Minimum conservation reference size ( \(\geq 27 \mathrm{~cm}\) ), whiting now mature well below this MCRS.
- MSY Brrigger \(=16300\) t
- \(\mathrm{F}_{\text {MSY }}=0.22\)
- \(\mathrm{B}_{\text {lim }}=10000 \mathrm{t}\)
- \(\mathrm{B}_{\mathrm{pa}}=16300 \mathrm{t}\)
- \(\mathrm{F}_{\text {lim }}=0.37\)
- \(\mathrm{F}_{\mathrm{pa}}=0.22\)
- SSB \(_{\text {MGT }}=\) Not applicable
- F \(_{\text {MGT }}=\) Not applicable

\section*{8. General Comments:}
- The report is very systematic, and well written. This stock had a benchmark assessment in 2017 (ICES 2017) and the assessment follows the stock annex. The input fisheries data and data issues are generally well documented and discussed. The WG does a good job discussing the sources of uncertainty and as well as further discussion of data issues.
- The RG found that the majority of catches have been discarded for the past couple of decades and discards information remains very imprecise. This statistical bias of data will have influence on the stock assessment. The RG suggests considering of a sensitive analysis to evaluate the impact of a variety of mis-specifications in fishery data on stock assessment results.
- The RG noticed that there are five surveys monitoring the Whiting, but the RG noticed that the assessment only includes three surveys as input data. The RG suggests providing the reasons for excluding other surveys data in the assessment report or stock annex. For the survey index that does not have a good fit in the model, the RG suggests providing more explanations in the assessment.
- Regarding Figure 1, what maximum effort of the main fleets can be expected under management based on \(\mathrm{F}_{\text {MSY }}\) (ranges) for the target stocks, and has the stock experienced similar levels of fishing effort before? The RG recommends clarifying this question.
- The RG noted that M is fixed to be a constant rate of 0.2 . However, the RG suggests adding more biological clarification to justify the choice of fixed M . In addition, the RG recommends (1) comparing the rate employed to the rates used for similar species in the same area or the same species in different areas, and (2) conducting a structured sensitivity
analysis to evaluate impacts of uncertainty in time-varying and age-varying \(M\) in the future assessments.
- The RG noticed that there is no explanation about model parameter steepness (h) in this stock assessment. Because the value of \(h\) may have a strong effect on the ASAP model outputs, The RG suggests clarifying the determination of the value.
- In this assessment, the WG conducted a retrospective analysis by removing 5 years of data. The RG suggests removing more years to cover the lifespan of the species for the retrospective analysis.
-The RG noticed that the report didn't include the comparison of the ASAP assessment and XSA run. The RG also suggests consideration of an XSA run with the ASAP run about uncertainty bounds comparison.

\section*{9. Technical Comments:}
- "The observed and predicted catches are shown in Figure 40.7. Fit to the overall catch is reasonably good. There is some deviation in the early to mid-1990's. This is most likely due to the introduction of the survey data into the assessment model." Although fit of the overall catch is reasonably good, there is a great discrepancy between observed and predicted catch in the early to mid-1990's (Figure 40.7). The RG suggests checking the settings for the early years and conducting diagnostics on this issue.


Figure 40.7 Whiting 7.a. Observed and Predicted Catches
- "The observed and predicted index cpue values are shown in Figure 40.8. There is poor fit to the Northern Irish ground fish survey indices in the first half of the series but it improves in recent years." The RG noticed that there is a discrepancy between the Northern Irish ground fish survey indices Q1 and Q4 during 1980-2005 year (Figure 40.8), the RG suggests the WG to make various assumptions of the quality of the input data by testing different levels of CVs, as the RG could not find any discussion of this issue in the report.


Figure 40.8 Whiting 7.a. Observed and Predicted index срие
- "Figure 40.9 shows the retrospective analysis. The predicted catch shows no obvious retrospective pattern, neither does the recruitment estimate. There is some deviation in the early part of the time series when the surveys were first introduced. However, recent estimates of SSB and \(F\) are consistent with no apparent bias." The RG appreciates that a retrospective analysis was provided for this stock, and used Mohn's Rho to quantify the uncertainty for this retrospective analysis. However, the Mohn's rho values for SSB (0.129) and recruitment ( 0.29 ) are still high. So the RG suggests the WG to adjust the estimated stock status using the estimated Mohn's rho value.


Figure 40.9 Whiting 7.a. Retrospective analysis of the final ASAP run with Mohn's Rho calculation.
-The RG also has the same question with the EG about ICES stock data category reference (ICES, 2018a).The link is correct but we don't know where the advice ended up in the end. The RG suggests please marking the page number or chapter in the Table 6.
- "Benchmarked" is misspelled in the whg.27.7a assessment report.
- Figure 40.8 The icon should be placed inside the chart.
- Figure 40.9 extra descriptions should be addressed for color code as there is no legend.
- Section 40.1- There is no caption on the map, and the coverage of the survey data should be provided.

\section*{10. Conclusions:}
- The assessment of haddock in Whiting (Merlangius merlangus) in Division 7.a appears well done and follows the stock annex. The RG suggests the stock assessment to be accepted with caveats.
- The RG suggests conducting exploratory runs to investigate the sensitivities of the results to changes in fishery input data, natural mortality values, and model parameter steepness (h).
- The RG recommends providing information on maximum effort of the main fleets can be expected under management based on \(\mathrm{F}_{\text {MSY }}\) (ranges).
- For the retrospective analysis, the RG suggests removing more years of data that cover the lifespan of the species and check the retrospective patterns.

\section*{39 Whiting (Merlanglus merlangus) in divisions 7.b-c and 7.e-k (Southern Celtic Seas and eastern English Channel)}

\section*{1. Assessment Type: Update}
2. Assessment: Accepted

\section*{3. Forecast:}
- Short term - Forecast for the next three years from 2017 (2018-2020); tuned with a single combined survey index.
- No medium or long term projections.
4. Assessment Method: Extended survivor analysis (XSA) with landings and discard data

\section*{5. Consistency:}
- The model (XSA) used the same settings as last year's assessment, using a truncated time-series from 1999-2017.
- Overall, the model estimates are reasonably consistent for ages \(1+\) (given whiting stocks are prone to yearly effects in survey catches).
- The internal consistency of the surveys was examined using pairwise scatterplots of log numbers-at-age. The index is reasonably consistent for older ages (Ages 1-5).
- Cohort and year effects were examined with mean log standardized plots.
- In the past three years, F experienced generally upward revisions and SSB generally downward revisions.

\section*{6. Stock Status:}
- Whiting is a category 1 stock with considerable data inputs on a yearly basis.
- Whiting is prone to year effects in survey catches. Recruitment is episodic. Strong recruitments occurred in 2009 and 2013. Overall, catches and SSB tend to fluctuate depending on year-class strength.
- Fishing mortality has increased since 2012 and is now assessed to be above Fmsy.
- In the current time-series, SSB displayed a peak biomass in 2012 (due to strong recruitment in 2009) and again in 2015.
- SSB is now estimated to be below Blim, the precautionary limits for this stock.
- Mean weight-at-age appears to have declined during the period of recent high fishing effort and landings between 2005-2008.
- Different gear types are used in this fishery, including trawl nets (mostly otter trawlers, also including bottom, beam, pelagic, and midwater trawls) and seines nets.
- Whiting stocks are some of the least limiting stocks for most fleets in Celtic Sea mixed fisheries. Still, discards are a prominent feature in this fishery.

\section*{7. Management and Biological Reference Points:}
- The management of this stock follows the multiannual management plan (MAP) for the Western Waters by the European Parliament and the Council in 2019.
- The whiting stocks have been managed through a TAC system covering 7.b-k, which encompasses those considered in this assessment area, 7.b,c,e-k.
- A TAC of 19184 t is set for 2019, down from 22213 t set for 2018. All are allocated within the EU.
- In 2016, the catch total ( 22457 t ) approached the TAC level ( 22778 t ). In 2017 and 2018, TAC ( 27500 t for 2017, 22213 t for 2018) has far exceeded catch levels ( 17780 t for 2017, 12625 t for 2018).
- Since 2017, a landings obligation has been applied to these fisheries in accordance with EU laws. Thus, all catches of whiting in the Celtic Sea and Western Channel must be landed. ICES estimates of discard rates indeed have been reduced to \(15 \%\) in 2018 from \(26-32 \%\) from the three preceding years.
- Minimum Conservation Reference Size for whiting at 27 cm . The report does note, however, that a significant portion of the unwanted catch is above this size.
- Current biological reference points:
\[
\begin{array}{ll}
\circ & \mathrm{B}_{\lim }=25000 \mathrm{t} \\
\circ & \mathrm{~B}_{\mathrm{pa}}=35000 \mathrm{t} \\
\circ & \mathrm{~F}_{\lim }=1.120 \\
\circ & \mathrm{~F}_{\mathrm{pa}}=0.900 \\
\circ & \mathrm{~F}_{\max }=\mathrm{N} / \mathrm{A} \\
\circ & \mathrm{~F} 0.1=\mathrm{N} / \mathrm{A} \\
\circ & \mathrm{~F}_{\mathrm{MSY}}=0.524 \\
\circ & \text { MSY } \mathrm{B}_{\text {trigger }}=35000 \mathrm{t}
\end{array}
\]

\section*{8. General Comments:}
- This report is concise and follows the stock annex.
- Data used and potential data issues are discussed in depth. There are fairly detailed discussions of sources of uncertainties.
- The report uses appropriate and extensive data for XSA models, such as catch-at-age data and landings data.
- The report remarks that bycatch and discards are a prominent feature in this fishery. In this case, a small paragraph on post-release mortality rate would be helpful to better understand this fishery and model projections.
- The RG recommends investigating the effects of environmental conditions on recruitment to better understand what causes high recruitment events.
- In this figure below (Figure 3 from the assessment), it seems that the total catch has experienced an inconsistent but prolonged decline from the beginning of the survey data in 1999. This assessment should discuss the reasons for this observation.


\section*{9. Technical Comments:}
- Overall, the text should be double-checked for grammar and clarity issues.
- Overall, the report will benefit from more graphic representations of extensive data tables to make explicit discernible patterns, such as year effects.
- Table 4 detailing catch and landings data is formatted in a confusing way, lacking column headings and row headings.
- Table 9, which details the output results from XSA diagnostics, should be made clearer with better formatting (.g. the yearly fishing mortality data, the XSA population number, and all other data-heavy charts should be put into a table format for better alignment).
- This report will benefit from a more in-depth discussion of the use of survey data to "tune" the model.
- Each panel in Figure 1 (Irish landings for the main gear types) are too small and appear illegible. Maybe include a legend to explain the \(x\)-axis values for the small graphs in each panel, as the average annual landings for each gear type are sometimes an order of magnitude different, making the graphic representation misleading.
- In Figure 2, the figure under OTB, the panel that shows UK(England) is squished and the y -axis values overlaps and are illegible.
- Overall, the graphs and figures included for data representation are well labeled and clear.

\section*{10. Conclusions:}
- The assessment of whiting appears well done and indicates no large retrospective errors.
- The assessment uses an appropriate model (XSA) and is supported by precise, extensive fishery data.
- The RG accepts the assessment, although more details about this fishery (biology, ecosystems, fishery) could enrich this report and inform future revisions.
- The RG also recommends investigating the effects of environmental conditions on recruitment to better understand what causes high recruitment events.

\section*{Pok.27.7-10 Saithe in ICES subareas 7-10}
1. Assessment Type: New assessment - never been assessed
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- Short-term projection - N/A
- Mid-term projection - N/A
- Long-term projection - N/A

\section*{4. Assessment Method:}
- Depletion-Corrected Average Catch (DCAC).
- Model Input: commercial catch data - time series of official landings reported.
5. Consistency:
- Saith in subareas 7-10 is considered a new stock to the WGCSE and has never been assessed.
- The average DCAC value was used to provide advice for 2019. The DCAC method estimated a yield likely to be sustainable utilizing the same model settings applied for Pollack in subarea 7.
- The average DCAC value was not used to provide advice for 2020. A TAC of no more than 582 tonnes was suggested for years 2020, 2021, and 2022 using the precautionary approach. The precautionary approach applies a \(20 \%\) buffer to the average landings from years 2016-2018.

\section*{6. Stock Status:}
- Current level of catch is well under the maximum sustainable yield and \(95 \%\) confidence intervals derived from the DCAC assessment method applied to commercial landings data since 1986.
- Insufficient information was provided to evaluate trends in saithe. In general, commercial catches have declined over time with the lowest values reported both in 2008 and 2018 and the largest values reported in 1986 at the beginning of the data set.
- 2018 values of landings reported 496 tonnes, a decrease of \(33.8 \%\) when compared to landings reported in 2017.

\section*{7. Management and Biological Reference Points:}
- No formal management plan has currently been established for saithe in subareas 710.
- Reference points have not been established as they cannot be estimated using the current assessment method.
- Total TAC of 3,176 tonnes was set for 2019. The precautionary approach suggests a TAC of no more than 582 tonnes in years 2020, 2021, and 2022.

\section*{8. General Comments:}
- The WG report is well written. Data and model limitations are well documented. No stock annex was provided.
- "The DCAC-method was applied during WGCSE 2019 with the same model settings as was applied to Pollack in subarea 7 due to the similarities of landings data". DCAC method settings should be further justified in the WG report.
- The RG suggests exploring the addition of discard rate utilizing data provided from the Irish fleet.
- The DCAC method is solely based on catch data and assumes accurate catch information. Discard and potential observation errors associated with the catch data for the saithe fishery should be considered when using this method.
- Explanation of why the DCAC method was not used to provide advice for 2020 is missing.
- The RG agrees with the WG that different assessment models should be explored over the coming year. The quality of supplementary biological or survey data that is available should be evaluated. All data gaps should be identified for potential models that are explored in the future.

\section*{9. Technical Comments:}
- Scientific name (Pollachius virens) belongs in the title or beginning of the advice document.
- Table X.1: caption should indicate units of landings officially reported to ICES. Meaning of asterisks located next to year 2017 and 2018 should be defined.
- Units should be provided for the figure displaying TAC values by country. This figure is embedded in the text under section "Management applicable to 2019".

\section*{10. Conclusions:}
- The RG suggests that the report be accepted with caveats.
- Saithe in ICES subareas 7-10 are classified as a category 4.1.2. stock which is a datalimited species. The data-limited method DCAC was explored this year to attempt to provide advice about the TAC. The performance of the DCAC is subject to the quality of catch data and accuracy of the pre-specified parameters and stock status (e.g. Fmsy.M, depletion level). The quality of data was not well discussed in the report. The setting of relevant parameters and stock status lacks justification. The sensitivity analysis the WG conducted was not sufficient to address the uncertainty
issues. The RG agrees with the WG that a stock assessment model should be explored since there are some available data other than catch. Before any model being applied for this stock, the RG recommends that the data gap be identified and the data quality evaluated.

\section*{Her 27.6a7bc: Herring (Clupea harengus) in divisions 6.a (combined) and 7.b-c (West of Scotland, West of Ireland)}
1. Assessment Type: update
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- No projections were carried out in 2019.
4. Assessment Method: Multi-fleet implementation of the State-space Assessment Model (Multi Fleet SAM) was used.

\section*{5. Consistency:}
- The model was changed to a multi fleet SAM assessment with data from 6.aN and 6.aS 7.bc treated separately.
- Improvements were made to the input data. The IBTS data series was recalculated using the delta GAM method and the acoustic surveys were combined into a single tuning index.
- The model remains sensitive to assumptions on age-dependent catchabilities, lack of information on recruitment and the abundance of fish of younger ages.
- The changes applied to the assessment following the interbenchmark in early 2019 have improved the results from the retrospective analysis but the bias is still present.
- The assessment in 2019 shows a very different perception of stock status with the 2018 assessment. The SSB has been significantly revised downwards and F has been revised upwards. Recruitment has also been revised downwards. SSB and recruitment are at very low levels with decreases in \(F\) evident in recent years.

\section*{6. Stock Status:}
- The assessment provides information for the combined stocks.
- Fishing mortality has been reduced since the introduction of zero catch advice and in line with the monitoring TAC in 2016
- Catch has been declining from 27123 tonnes in 2014. Recent catches have been amongst the lowest in the time-series. Total catch in 2018 was 5558 tonnes while the fishery has been restricted to a monitoring fishery with a combined TAC of 5800 tonnes since 2016.
- SSB has decreased steadily since 2003. The Malin Shelf herring estimate of SSB for 2018 is 159000 tonnes and 925 million individuals, a slight increase compared to the 145 000 tonnes and 798 million herring estimate in 2017. However, the estimate SSB for 2018 is still very low in the time-series.
- Recruitment has been declining from 2,602,809 tonnes since 2000 and is predicted to be at their lowest level in the assessment period (230,732 t for 2018). Recruitment has been low with no big cohorts evident in recent years.

\section*{7. Management and Biological Reference Points:}
- The managers are advised to ensure that any exploitation pattern imposed in this area will not overexploit the smaller and more vulnerable stock.
- No MSY and PA reference points were provided.
- ICES advises that when the precautionary approach is applied, there should be a zero catch in 2020.
- The 2019 TAC is agreed as 5800 tonnes.

\section*{8. General Comments:}
- The assessment of Herring in divisions 6.a (combined) and 7.b-c (West of Scotland, West of Ireland) appears well done, and the WG report is well written, following the procedure agreed by the recent interbenchmark (IBP6a7bc, ICES 2019). The data and data issues are well documented and the updated assessment provides the best statistical fit to the input data. The WG did extensive work for providing the sources of uncertainty.
- The WG thinks that the new model remains sensitive to assumptions on age-dependent catchabilities, and lack of information on recruitment and the abundance of fish of younger ages.
- The RG realizes the extensive work have been done for the sensitivity analysis, and suggests providing more information about the methodology, results and discussion of the reasons and solutions for the sensitivity issues.
- Specifically, the model is sensitive to age-dependent catchabilities while the pattern of increasing catchability with age cannot be explained. The catchability of one index WOS_MSHAS is very high. The RG suggests the WG conducting sensitivity analysis to test more assumptions for the catchability (e.g. logistic, random walk, and exponential pattern).
- The age structure varies from different data sources. The RG suggests that the WG could further explore which index is more representative and could weighting the data sources with consideration for their variance (Figure 4.6.6).
- The updated assessment provides the best statistical fit to the input data, but the assessment still has a strong retrospective bias, especially for fishing mortality. The Mohn's Rho on 5 -year peels is -23 . The RG realizes that the updated assessment following the interbenchmark has improved the retrospective but bias is still present.
- The RG recommends briefly listing the changes to assessment model and the comparison of the two retrospective biases, which could contribute to explore the explanation of bias and to further improve the assessment.
\(\circ\) Given the different standard bias for SSB, F and R, the RG recommends adding a "relative" difference compared to the terminal year among retrospective peels.
- Since herring have a maximum lifespan of 10 years, the WG is suggested to remove more years of data during the retrospective analysis.
- Since the retrospective error has a strong pattern and the magnitude of error is big, the final estimates of this assessment may need to be adjusted with the estimated bias. Also, the RG recommends that simpler models (such as surplus production and VPA) could be applied to double check the results.
- The assessment in 2019 shows a very different perception of stock status with the 2018 assessment, especially for SSB, F and R after 1991, while the patterns of SSB, F and R before 1991 are much similar. The RG suggests more discussion about the different patterns and different settings between the two assessments.
- In this WG report, no short-term projections were carried out in 2019. Since the monitoring TAC taken in 2016, there are a little recovery for this stock, such as the increasing SSB and decreased F. Given the differentiating pattern of this stock and its continuing strict TAC management strategy, the RG suggests conducting a simple shortterm projection following the Stock Annex, in order to explore the performance of this strategy and to predict the population dynamics in the next short term.
- Since some input data are not available for early years, these data are assumed to be the same with the first year of available data. For example, the proportions of maturity before 1991 are assumed to be the same with 1991, while the weight-at-age in stock before 1991, and weight-at-age in catch for stock \(6 . a N\) before 1984, the catch-at-age for stock \(6 . a S\), the 7.b-c before 1983. The RG suggests conducting a sensitivity analysis for quantifying the impact of these unavailable data in early years on the assessment results, such as using average values among the adjacent 2 or 3 years.

\section*{9. Technical Comments:}
- The RG suggests providing the references list after this section 4, which will improve understanding of this stock assessment report.
- Section 4.2- The 2013 year class (age 4-wr) dominates both in the catches and the acoustic survey in 2018. For the acoustic survey, the 2010 year class (age 1-wr) is the most dominated cohort for both abundance and biomass. The RG recommends reflecting this issue.
- Section 4.10- There is anecdotal evidence that the stocks are not the same size and managers are ad-vised to ensure that any exploitation pattern imposed in this area ensures that the smaller, more vulnerable, stock is not over-exploited. This sentence can revise to be more concise and clarified.
- Tables need to be listed in descending order according to table number. Some tables cited are not available in this WG report, such as Tables 5.1.2, 5.2.4 and 5.2.5.
- Does the Fbar in Tables 4.6.9 and 4.6.12, and Figure 4.6.10, means fishing mortality, same with F in the text and other tables and figures? Clarification may need for this consistent issue.
- Table 4.3.1.2- More information should be provided in the caption. The data source (Acoustic surveys) can be stated to be consistent with Table 4.3.1.3. It's better to add
decimals for the biomass of Immature, Mature and Total, in order to clear make Total=Immature +Mature.
- Table 4.3.1.3- The RG suggests using consistent Age (Rings) with Table 4.3.1.2, which is to add Age 0 and to replace Age 9 with Age 9+.
- Table 4.6.1a- The RG suggests using integer to represent individual numbers and being consistent format with table 4.6.1b.
- Table 4.6.5- The maturity in 2007 is not reasonable. The RG suggests adding some explanation for this, or to test the sensitivity of the original value, average value of 2006 and 2008, etc.
- Fractions of harvest and nature mortality before spawning in Tables 4.6.6 and 4.6.7 are 0.67 for all ages and all years. The RG recommends providing more information of this fixed value.
- Table 4.6.8. SURVEY INDICES are duplicate with Table 4.3.1.3. However, the units million and number in Table 4.6.8 and 4.3.1.3, respectively, make values inconsistent.
\(\bullet\) Figures 4.6.1~5- Is there different meaning for black and red bubbles? It's better to clarify this issue in the caption.
- Figure 4.6.6- The x axis labels are not shown completely for WOS-MSHAS. It's better to indicate that in x axis labels the numbers after data sources represent age (wr). Alternative order in the figure is to list the observation variances by the data source.
- Figure 4.6.7- The ages after the data sources are overlapped and unavailable to clarify.
- Figure 4.6.10- The uncertainties for key parameters (SSB, Fbar, and Rec) are extremely high in recent years. The RG suggests providing more discussion about this issue, not only for recruitment.

\section*{10. Conclusions:}
- The WG did an excellent job in their assessment of herring in divisions 6.a (combined) and 7.b-c (west of Scotland, west of Ireland) by providing the best statistical fit to the input data. However, the assessment still has a strong retrospective bias, especially for fishing mortality.
- Given the updated model shows the best fit to the data, and high uncertainties of key parameters and sensitivity to many assumptions, the RG suggests the stock assessment to be accepted as long as following concerns are addressed: a) providing more information about the existing sensitivity analysis and different settings with the previous assessment; b) alternative sensitivity analysis for the unavailable biology data and more assumptions for catchability as discussed in the general comments.

\section*{Herring in the Celtic Sea (Clupea harengus) in Divisions 7.a South of \(5^{\circ}{ }^{\circ} 30^{\prime}\) N, 7.g, 7.h, 7.j and 7.k (Irish Sea, Celtic Sea, and southwest of Ireland)}
1. Assessment Type: Update, benchmarked in 2015, inter-benchmarked in 2018
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- Short term projection - A deterministic version was carried out following the procedure of 2014 benchmark.
- Long-term projection -Simulations were carried out to evaluate the long-term management plan. The plan was proved not precautionary and thusly not used to give advice. Development of rebuilding plan is in process.
4. Assessment Method: Age Structured Assessment Program (ASAP)

\section*{5. Consistency:}
- The model configuration is the same as the 2018 inter Benchmark,
- The 2019 assessment was supported by the Celtic Sea herring acoustic survey (CSHAS). The survey had two spatial components and five replicates. Survey data from 2004 and 2017 were excluded in the 2019 assessment.
- The 2018 survey showed the lowest estimate in the current time series because herrings have been moving below the detectable depth of acoustic survey since 2014.
- The mean weight-at-age has been declining since 1990 and reached almost historical low in recent years. The shift towards late spawning in the region was reported by both scientists and fisherman. These were caused by adverse changes in the environment. However, herring in the region showed no sign of regime shift.
- The 2019 ASAP output indicated a lower selection-at-age for 9-wr (winter rings) fish. This might be caused by its low abundance in catch.
- The 2019 ASAP output indicated extremely high uncertainty of SSB, Fbar, and recruitment.
- The 2019 ASAP output indicated more pessimistic outlook for the stock with downward revision of SSB and upward revision of F.
- The long-term plan was reviewed and rejected in 2018. A new rebuilding plan is being developed.

\section*{6. Stock Status:}
- The 2018 SSB estimate was 22977t, which was below \(\mathrm{B}_{\mathrm{pa}}\) and \(\mathrm{B}_{\text {lim. }}\). SSB has been declining since 2011.
- The \(2018 \mathrm{~F}_{\text {bar }}\) estimate was 0.33 , which above \(\mathrm{F}_{\mathrm{pa}}\) and \(\mathrm{F}_{\mathrm{MSY}}\) and just below \(\mathrm{F}_{\text {lim }}\). Although it has decreased from 0.64 (2017).
- Total catch in 2018 was \(4418 t\) while the TAC was 10127 t.
- Recruitment of recent years has fallen below the long-term average since 2013.

\section*{7. Management and Biological Reference Points:}
- The closure (with limited fishing allowed) of the Subdivision 7.a.S was introduced since 2007-2008 to protect first time spawners. Currently only vessels shorter than 50 feet in registered length are permitted to fish in this area. A maximum catch limitation of \(11 \%\) of the Irish quota is allocated to this fishery.
- No long-term management plan is currently adopted. The previous one was proved risky while the new rebuilding plan is still being developed.
- Reference points were re-estimated in 2018 under both MSY approach and Precautionary approach frameworks.
- TAC for 2019 is 4742 t (based on the ICES MSY approach).
- Current reference points under MSY approach framework:
- \(\mathrm{F}_{\mathrm{MSY}}=0.26\)
- MSY B \(_{\text {trigger }}=54000 \mathrm{t}\)
- Current reference points under Precautionary approach framework:
- \(B_{\lim }=34000 \mathrm{t}\)
- \(\mathrm{B}_{\mathrm{pa}}=54000 \mathrm{t}\)
- \(\mathrm{F}_{\text {lim }}=0.45\)
- \(\mathrm{F}_{\mathrm{pa}}=0.27\)

\section*{8. General Comments:}
- The WG report is well written and follows the stock annex. The input data and associated issues are well documented and discussed in depth. The WG extensively evaluated the quality of the assessment from multiple aspects.
- According to Figure 6.6.1.9, CVs of key parameters estimations are historically high. The underlying uncertainty is considerable.
- The WG comprehensively introduced the design of Celtic Sea herring acoustic survey (CSHAS). The RG agree with the WG that incorporating the adaptive surveys will improve the stock assessment. However, the RG still has several concerns:
- The survey data was used following the statement from stock annex: "The Celtic Sea Herring Assessment is tuned using a single acoustic survey." The RG is not sure how did WG standardized the data sets from two different components into one. Were they treated as equally weighted?
- The RG would like to know if the level of variation from the two survey components are comparable. This is unfortunately not available in the stock annex or the report.
- According to the Table 1 from stock annex, the survey design was not consistent for the considered time-series. The RG is not sure if the consistency of survey design was violated in this case.
- According to the Table 6.6.1.2, the survey CV was assumed constant for the whole time series. The RG is not sure how robust is this assumption given the change in survey design.
- The WG conducted short-term projections with deterministic simulation. However, given the considerable uncertainty associated with the estimates of recruitment and SSB in recent years, the RG would suggest conducting a stochastic simulation to incorporate these uncertainty.
- The WG discussed the changes in the environment that might influence the wellbeing of the stock. Building on this, the RG would like to know how these changes would directly or indirectly bias the stock assessment and survey. Were these changes contributing to the recently high uncertainty in assessment?

\section*{9. Technical Comments:}
- "NASC distribution plots from the broad-scale survey are presented in Figure 6.3.1.2 and from the adaptive mini survey in Figure 6.3.1.3." The RG suggests presenting the full spelling the abbreviation NASC since it is technical.
- "Weights in the catch and in the stock at spawning time have shown fluctuations over time (Figures 6.4.4.1 and 6.4.1.2)......" The first cited figure should be 6.4.1.1
- Table 6.6.1.1, unit is missing for stock and catch weight-at-age.
- Table 6.6.1.2, the RG is not sure if the presented age-specific values were standardized. Unit is also missing here.
- Table 6.6.1.3, "sample size" should be "No. of herring...." Instead of "No of herring....". Also, why Fbar was calculated for ages 2-5? The first fully selected age group is three according to Figure 6.6.1.4.
- Table 6.6.1.4, units are missing in this table. The full spelling of TSB should be noted as well.
- Table 6.7.1.1, unit is missing for " N ".
- Table 6.7.1.2, units are missing in this table.
- Figure 6.2.1.1, y axis labels and units are missing.
- Figure 6.3.1.4, the label for x and y axis are the same here. Should they be "observed" and "predicted"?
- Figure 6.4.1.1, unit is missing in this figure.
- Figure 6.4.1.2, unit is missing in this figure.
- Figure 6.6.1.2, x axis label should be "Year", y axis label should be "Value in tons".
- Figure 6.6.1.8, there is no left or right panel in this figure.
- Figure 6.10.1, legend is missing for this figure.

\section*{10. Conclusions:}
- The RG suggest this assessment should be accepted with caveats.
- The assessment of herring in the Celtic Sea in Divisions 7.a South of \(52^{\circ} 30^{\prime} \mathrm{N}\), 7.g, 7.h, 7.j, and 7.k appears well done and flawless in terms of the assessment process.
- However, there are also some issues regarding the use of survey data need to be further addressed. Furthermore, extremely high uncertainty was exhibited by estimates of key parameters. The RG suggests this should be considered in the future assessment and management using stochastic projection to for precautionary consideration.

\section*{her.27.nirs: Herring (Clupea harengus) in Division 7.a North (Irish Sea)}
1. Assessment Type: Update, Benchmarked in 2017
2. Assessment: Accept with caveats

\section*{3. Forecast:}
- Short-term projection- Carried out by a FLSAM using data from 1980-2018 to predict 2019-2020. SSB is expected to be well above MSY \(\mathrm{B}_{\text {Trigger }}\) in 2019-2020, but will likely decrease if fished at \(\mathrm{F}_{\text {MSY }}\).
- No medium or long term projection was carried out.
4. Assessment Method: State-Space Age-Structured Model, in the FLR environment (FLSAM) was used.
- Model Configurations
- SSB-R relationship: cohort back-tracking from older ages.
- Natural mortality: Values from the North Sea, not the Irish Sea. Age specific mortalities were defined and held constant over the time series.
- Major Input Data
- Catch Survey (fisheries dependent): numbers-at-age, numbers-at-length, total landings, weights-at-age, abundance index
- Acoustic Surveys (fisheries independent): SSB estimates, length frequencies
- SSB Survey (fisheries independent): SSB index, recruitment
5. Consistency:
- The assessment model did not change in 2018 after the stock was benchmarked in 2017 but the following changes occurred following the benchmark.
- Future analyses will use landings data only dating back to 1980 instead of 1961 due to data quality concerns (justification in annex).
- A random walk assumption on recruitment was removed and replaced with estimations of recruitment from cohort back-tracking from older ages.
- The stock is considered mixed, and an SSB survey index using acoustic methods was implemented in place of a larval survey to eliminate some age and mixing issues. More weight was given to the SSB survey in an effort to give a more balanced model not simply based on catch.
- There was a marked increase in in landings made by Irish vessels in 2018 comprising \(19 \%\) of all landings compared to an average of \(2 \%\) in the preceding three years.

\section*{6. Stock Status:}
- In 2018 SSB and the highest F value (experienced by age 2 and 4 fish- \(\mathrm{F}_{\text {bar }}\) is not provided) were 39,997 tonnes and 0.19 respectively. SSB was above MSY \(\mathrm{B}_{\text {Trigger }}\) and \(F\) was below \(\mathrm{F}_{\text {MSY }}\).
- Trends show an increase in SSB and recruitment since the mid-2000s, stabilizing in recent years (although uncertain).
- Landings in 2018 were estimated as 6804 tonnes, while TAC was 6896 tonnes
- Mean weights-at-age have shown a general downward trend.

\section*{7. Management and Biological Reference Points:}
- TAC for 2019 was not defined, but 6896 tonnes or less was advised by ACOM.
- Year-round closures occur along the east coast of Ireland and within 12 nautical miles of the west coast of Britain (Scotland, England and Wales). The traditional gillnet fishery on the Mourne herring is exempted from these closures.
- Seasonal closures of the Douglas Bank spawning ground, to the east of the Isle of Man occur from September \(21^{\text {st }}\) to November \(15^{\text {th }}\). Boats from the Republic of Ireland are also not permitted to fish east of the Isle of Man.
- No current long-term management plan is in place for this stock. The RG agrees with the WG in supporting the development of a long-term management plan.
- New Reference Points for 2018:
\begin{tabular}{lll}
\hline \multicolumn{1}{c}{ Reference point } & Value & Technical Basis \\
\hline \(\mathrm{B}_{\text {lim }}\) & 8500 t. & \begin{tabular}{l} 
Set to lowest SSB that \\
generated above average \\
recruitment
\end{tabular} \\
\hline \(\mathrm{B}_{\mathrm{pa}}\) & 11800 t & \(\mathrm{B}_{\text {lim }} * \mathrm{e}^{1.645 \sigma}, \sigma=0.201\) \\
\hline \(\mathrm{MSY} \mathrm{B}_{\text {Trigger }}\) & 11800 t & \(\mathrm{B}_{\mathrm{pa}}\) \\
\hline \(\mathrm{F}_{\text {MSY }}\) & 0.27 & \begin{tabular}{l} 
Upper and lower bounds \\
(median), \\
giving at least \(95 \%\) of the
\end{tabular} \\
& \begin{tabular}{l}
0.35 (upper), \\
maximum yield
\end{tabular} \\
& 0.20 (lower)
\end{tabular}

\section*{8. General Comments:}
- The report is well written, and the limitations and data issues are discussed in detail. This report does a good job of providing the sources of uncertainty.
- "Currently, the model doesn't have the structure to specifically deal with the emigration of small herring from other stocks". There is known mixing of young herring from the Celtic Sea into this area, and the annex for South Celtic Sea herring suggests that there is a difference in growth rates between these groups that is likely to impact recruitment. The mixing issue has not been addressed since the 2012 benchmark, and The RG suggests new mixing studies be performed and that an effort is made to account for this difference in growth when determining weight-at-age, particularly in younger fish when mixing occurs more frequently. The RG also suggests updating the annex to include information about growth rate discrepancies between Irish and Celtic herring.
- There is strong uncertainty surrounding catchability of herring in the SSB acoustic surveys. Catchability values of \(1,2.5\) and 3 are mentioned in the report, but it is not clear which value was used in the analysis nor the justification for using this value.

The RG appreciates the sensitivity analysis performed on catchability, and suggests that the methods and the results of the sensitivity analysis also be included in the report. The RG also recommends that the WG revisit the results of this analysis since, "the reviewers could not reach a consensus and posed that HAWG is the best place to propose a final assessment model" and come to some consensus on how to use the results to better the model.
- "Mohn's rho was reduced from 13.3 to \(9 \%\) under shortened time series", after removing years before 1980, but the WG is not clear what this value of Mohn's row is referring to (SSB, Fbar, or recruitment). The RG suggests providing Mohn's rho values for all 3 parameter's retrospective analysis. The RG also recommends that the WG include methods for the regressions analysis (number of peels etc.) and a description of results.
- Natural mortality data is taken from the North Sea due to a lack of understanding of natural mortality in the Irish Sea, and uses constant age-dependent M values from year to year. The RG recommends calculating the natural mortality of the Irish Sea Stock using life history parameters of the Irish Sea stock, and justifying the use of constant M over time.
- The SSB acoustic survey aims to better understand the spatial and temporal variability of herring, but optimal timing of the survey has not been determined due to interannual variation in migration patterns. The RG suggests evaluating the accuracy of this data and exploring different methods to gain better insight into movements of herring.
- Due to herring's sensitivity to changing environments, the RG recommends gathering information on how SSB and recruitment are affected by changes in temperature.

\section*{9. Technical Comments:}
- The scientific name of the species, "Clupea harengus" should be included in the report.
- The RG recommends including a map of stock boundaries and study areas into the report.
- Section 7.2.2 - "The 2017 benchmark concluded to conduct future assessments only "data" back to 1980 " should read, "The 2017 benchmark concluded to conduct future assessments only "dating" back to 1980"
- Section 7.3.2 - "The 2012 benchmark (ICES WKPELA 2012) also suggested that the survey series could be used to fine tune the main survey used as the tuning fleet in the assessment The survey used a stratified design similar to the AC(7.aN." should have a period after the word, "assessment" and should have closing parentheses in the last section reading, " \(A C(7 . a N)\)."
- Section 7.3.2 - "The results of the survey "is" reported in the WGIPS 2018 report (ICES 2018)" should read, "The results of the survey "are" reported in the WGIPS 2018 report (ICES 2018)"
- Section 7.6.1 - "This was completed post-benchmark, however, the reviewers could not reach a consensus and proposed that HAWG is best place to propose a final assessment model" should read, "This was completed post-benchmark, however, the
reviewers could not reach a consensus and proposed that HAWG is "the" best place to propose a final assessment model"
- Section 7.6.1 - The additional period should be deleted.
- Table 7.6.3.3. Year 2018 should be in Times New Roman Font.
- Table 7.6.3.4. Natural mortality values for each age can be described in the report, and the model inputs can be placed in the annex.
- Table 7.6.3.6. The fraction of harvest before spawning can be described in the report, and the model inputs could be placed in the annex.
- Table 7.6.3.7. The fraction of natural mortality before spawning can be described in the report, and the model inputs can be placed in the annex.
- Tables 7.6.3.12-7.6.3.25. Times New Roman Font.
- Table 7.6.3.12. Table should be better centered on the page. Units needed.
- Table 7.6.3.14. Data from 1983-1986, 1990-1993, 1997-2000, 2003-2007, 20112016, and 2018 are missing from the data set. Age 8 of years 2001-2003 should not have bullet points. Units needed.
- Tables 7.6.3.15 \& 7.6.3.16. Data is missing.
- Table 7.6.3.18. Units needed.
- Table 7.6.3.19. Format year 2000 data next to year 1999 data.
- Table 7.6.3.20. Units needed.
- Table 7.6.3.22. Units needed.
- Table 7.6.3.26. Not a table, just a single number. Put this number in the report or in the annex and delete.
- Table 7.7.1. No line at the bottom of page 559 or on the top of page 560.
- Table 7.7.2. Units needed.
- Figure 7.1.1. Label \(x\)-axis. Reformat figure caption. The figure should be moved so that the right border can be seen.
- Figure 7.3.2. Figure caption should be on the same page as figure.
- Figure 7.3.5. X -axis should be labeled. X -axis should have fewer increments, it is difficult to determine which tick marks belong to which year. Times New Roman Font.
- Figure 7.4.1. Each data point should not display the age at which it is referring to. The legend on the right should only show the age associated with a color. X-axis should start closer to the year of the first data point.
- Figures 7.6.1-7.6.16. Last word in each figure caption should have a space between "age" and the number, ex: "agel" should be "age l".
- Figure 7.6.18. X-axis should have more increments, one for every year. Include legend or a description in the caption of what each line represents.
- Figure 7.6.19. Reformat figure caption. Describe age units (years, rings etc).
- Figure 7.6.20. Reformat figure caption.
- Figure 7.6.21. Reformat figure caption.
- Figure 7.6.22. Reformat figure caption. SSB and Recruitment units needed.
- Figure 7.6.23. Reformat figure caption.
- Figure 7.6.24. Reformat figure caption. SSB and Recruitment units needed.
- Figure 7.6.25. Reformat figure caption. Units needed.

\section*{10. Conclusions:}
- The assessment of herring in division 7.a north (Irish Sea) is very well organized and did a great job identifying potential sources of uncertainty and bias in the sampling methods and model configuration. The current stock seems healthy and at little risk of biomass and fishing mortality levels being outside of acceptable biological limits in 2019 if the current trends continue. However, the RG agrees with the WG that a longterm management plan should be developed for herring in this region. The RG recommends continued research into uncertainty caused by spatial and temporal variability, stock mixing, natural mortality, and catchability assumptions to continue improving model fit. Description of the methods and discussion of results (including Mohn's rho values) of the retrospective analysis should be included in the report, and the RG recommends reviewing all figures and tables to ensure font, unit and figure caption consistency.

\section*{Spr.27.67a-cf-k - Sprat (Sprattus sprattus) in Subarea 6 and Divisions 7.a-c and 7.f-k}
1. Assessment Type: Update
- No analytical assessment used for this area.
- Stock Annex last updated 2013
2. Assessment: Accepted with Caveats.

\section*{3. Forecast:}
- Short-term projection: N/A
- Mid-term projection: N/A
- Long-term projection: N/A
4. Assessment Method: None
\(\bullet\) None used

\section*{5. Consistency:}
- Seven survey efforts are referenced, but none provide sufficient data for region wide analysis.
- Historical landings for Ireland may be inaccurate, but no scale of inaccuracy is suggested.

\section*{6. Stock Status:}
- There is no stock assessment for sprat in areas 7.a-c and 7.f-k.
- Biomass predicted by intermittent survey effort shows a highly variable stock with large inter-annual variations in abundance determined by recruitment variability.
- The state of this stock is currently unknown with insufficient data for determining status.
- Uncertainty in stock structure may cause localized depletion.

\section*{7. Management and Biological Reference Points:}
- There is no sprat specific management plan.
- Sprat fishing effort is not quota controlled outside of 7.d-e, and the relationship between 7.d-e, 7.a-c and 7.f-k population is unknown.
- Sprat fishing effort is only controlled by a herring bycatch ceiling limit and bycatch percentage limits as herring and sprat are typically fished together.

\section*{8. General Comments:}
- The RG admires the attention to ecological importance of this species and mention of seabird and marine mammal prey dynamics.
- There is biological data for area 6a but this data is not available electronically. Given the data deficiency of this species, access to this biological data may be useful for future assessments.
- Given the WG statement that there is uncertainty associated with Irish small vessel historical landings, the RG suggests including confidence intervals for these uncertain data.
- The ICES suggested catch limit of 2800 tonnes is given by referencing previous years, by applying a precautionary approach to the 2013 decision of 3500 tonnes. However there is no mention of how this initial 3500 tonnes limit was chosen. The RG suggests including information on how this limit was chosen.
- The Sprat Stock Annex has not been updated since 2013. Since the estimated biomass given in the HAWG report is several magnitudes larger than landings this species does not appear to be in peril, but survey data and biomass estimates have accumulated that may be worth updating in the Stock Annex.

\section*{9. Technical Comments:}
- The HAWG report figures 13.2.1, 13.2.2, 13.2.3, 13.2.4, 13.2.5, 13.2.7, and 13.3.1 are missing axis labels.
- The HAWG report Section 13.1.2 Division 6.a cites a nonexistent figure (10.2.1) for landings by nationality from 1972-1978.
- The HAWG report 13.2.1-13.2.7 plots are difficult to interpret, area plots appear visually confusing. The RG suggests exploring different presentation methods for this data.

\section*{10. Conclusions:}
- Based on the high variability in stock estimates for this species, and an unknown stock recruitment relationship, the RG supports the precautionary limit of 2800 tonnes for 2020 and 2021.
- The RG accepts this assessment under the following caveat.
- The RG suggests adding some context for the 20133500 tonnes limit decision the current limit is based on.

\section*{Spr.27.7de: Sprat (Sprattus sprattus) in divisions 7.d and 7.e (English Channel)}
1. Assessment Type: Update
2. Assessment: Accepted
3. Forecast: No forecast was carried out.
4. Assessment Method: Biomass trend

\section*{5. Consistency:}
-The PELTIC acoustic survey was expanded to cover the southern area of division 7.e in 2017 and was extended into division 7.d in 2018.
-The age classes of the samples collected during the PELTIC acoustic survey decreased from 6 classes ( 0 to 5) in 2017 to 4 classes ( 0 to 3 ) in 2018.
-The stock identity for sprat in the English Channel is still unclear.
-Discards occurred but were considered negligible during 2014-2018.
-The biomass trend estimated by the PELTIC acoustic survey has been used as the index of stock development since 2016.
-The sprat landings were not exceeding the TAC in recent years, but the difference is getting closer.

\section*{6. Stock Status:}
-The stock status relative to MSY and PA cannot be assessed because the reference points are unknown.
- The total landing has decreased from 2733 tonnes in 2017 to 2252 tonnes in 2018.
- The LPUE (kg/day) has decreased from 9457 tonnes in 2017 to 8373 tonnes in 2018.
-The biomass estimate in 2018 ( 17091 tonnes) is less than 2017 value ( 32751 tonnes) but is still twice of that of the lowest level of the time series.
-The harvest rate (the ratio between catches and the acoustic index) peaked in 2016 (34\%), and was low (around \(10 \%\) ) throughout the recent 5 years.

\section*{7. Management and Biological Reference Points:}
-The catch advice was based on the data limited method, "2 over 3" rule (the ratio between average biomass of the last 2 years and average biomass of the previous 3 years) which was used to determine whether the uncertainty cap should be applied.
-The ratio resulting from " 2 over 3 " rule was 0.47 and a \(20 \%\) uncertainty cap was applied.
-The agreed TAC was set equal to 3296 tonnes and 2637 tonnes for 2018 and 2019, respectively.
-The advice catch was set equal to 1883 tonnes and 1506 tonnes for 2019 and 2020, respectively.
-No biological reference points were included in this assessment.

\section*{8. General Comments:}
-The WG report was overall well written and followed the stock annex. The fisheryindependent survey information for small pelagic fish community was well documented at length. The WG did a good job updating the assessment for this year. The spatial distribution of the stock was well investigated.
-The RG is not sure whether the data of oceanographic condition was collected during the PELTIC acoustic survey and whether any study has been done to understand the relationships between environmental variables and species using the survey data. It was not introduced in either WG report or stock annex. The RG recommends that the WG study how environmental factors (such as temperature and salinity that were mentioned in section 12.1.2 "Landings") affecting the biomass fluctuations, spatial distribution and fishery for sprat.
-There was a great discrepancy in the age composition from samples between the PELTIC acoustic survey and the FSP acoustic survey, which may be caused by gear selectivity. Further comparison of the representativeness of the samples from two surveys is needed. A careful comparison would also help the future analysis that requires the age composition data, which depend on the gear selectivity.
-The current assessment is primarily based on survey indices and the current precautionary approach is based on an empirical method. The RG recommends the WG apply various analytical assessment methods (e.g. SPiCT with environmental variables) based on best available survey information.
-The RG agrees with the WG's suggestion that further investigations are required to identify the stock structure of sprat populations in the English Channel. The RG also agrees with the WG's concerns on the connection between the Western English Channel stock and the Bristol Channel, where large numbers of juveniles were found. More investigations, such as morphometric and genetic analyses, are required to resolve this uncertainty.

\section*{9. Technical Comments:}
-The WG stated the inter-annual variability in stock abundance for sprat was mainly driven by recruitment variability. The RG encourages WG to explore the effect of recruitment variability on large internal-annual fluctuations in stock biomass with the survey data.
-The stock annex states that "The trend in biomass from both the PELTIC acoustic survey and the CPUE index from IBTS Q4 survey will be monitored and compared to support the advice: in case of consistent signs of impaired biomass from both or from
either one of the two indices, a warning should be issued and additional measures should be taken", there is no description in the report.
-The RG notices the stock biomass in 2016 was very low and the total landings around 1980 was considerably high, however, the WG did not provide any explanation for these. The RG recommends the WG explain these historical variations in stock biomass.
-Table 12.1.1- The landings data in the table (since 2014) was slightly different between this year and last year. The different part was highlighted in yellow below. The RG recommends the WG clarify this data revision in the report.
Table 12.1.1 Sprat in 7.d-e. Landings of sprat, 1985-2018. (from HAWG Report 2019)
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline Country & Denmark & France & Netherlands & UK Eng+Wales+N.Irl. & \begin{tabular}{l}
UK \\
Scotland
\end{tabular} & Other & Total \\
\hline 2012 & 6 & 2 & 8 & 4458 & 0 & 0 & 4474 \\
\hline 2013 & 0 & 0 & 0 & 3793 & 0 & 0 & 3793 \\
\hline 2014 & 45 & 0 & 275 & 3338 & 0 & 0 & 3658 \\
\hline 2015 & 0 & 1 & 352 & 2659 & 0 & 0 & 3012 \\
\hline 2016 & 185 & 7 & 231 & 2867 & 0 & 49 & 3339 \\
\hline 2017 & 0 & 0 & 235 & 2498 & 0 & 0 & 2733 \\
\hline 2018 & 474 & 1 & 0 & 1776 & 0 & 0 & 2252 \\
\hline
\end{tabular}

Table 12.1.1 Sprat in 7.d-e. Landings of sprat, 1985-2017. (from HAWG Report 2018)
\begin{tabular}{|l|llllll|l|}
\hline Country & Denmark & France & Netherlands & Germany & \begin{tabular}{l} 
UK \\
Eng+Wales+N.Irl.
\end{tabular} & \begin{tabular}{l} 
UK \\
Scotland
\end{tabular} & Total \\
\hline 2012 & 6 & 2 & 8 & 0 & 4458 & 0 & \\
2013 & 0 & 0 & 0 & 0 & 3793 & 0 & 374 \\
\hline
\end{tabular}
\begin{tabular}{|l|lllllll|l|}
2014 & 45 & 0 & 275 & 0 & 3358 & 0 & 3678 \\
2015 & 0 & 1 & 346 & 0 & 2657 & 0 & 3003 \\
2016 & 185 & 7 & 231 & 49 & 2867 & 0 & 3339 \\
20 & 0.03 & 235 & 0 & 2498 & 0 & 2733 \\
\hline
\end{tabular}
-Table 12.6.1- The biomass estimates from the acoustic survey were slightly changed in report from 2018 to 2019. The RG recommends that the WG clarify this data revision in the report.
Table 12.6.1. Sprat in 7.d-e. Annual sprat biomass in ICES Subdivision 7.e (Source: Cefas annual pelagic acoustic survey). (from HAWG Report 2019)
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|}
\hline Survey & Area & Season & 2011 & 2012 & 2013 & 2014 & 2015 & 2016 & 2017 & 2018 \\
\hline Partial & Lyme Bay & Oct & 33861 & 24246 & 62040 & 67538 & 12212 & 6181 & 29996 & 15310 \\
\hline FSP & Lyme Bay* & Oct & 33861 & 27971 & & & & & & \\
\hline PELTIC & W Eng Ch & May & 85358 & & & & & & & \\
\hline PELTIC & W Eng Ch & Oct & & & 70680 & 85184 & 65219 & 9826 & 32751 & 17091 \\
\hline
\end{tabular}
* ICES rectangles 29E6, 30E6

Table 12.6.1. Sprat in 7.d-e. Annual sprat biomass in ICES Subdivision 7.e (Source: Cefas annual pelagic acoustic survey). (from HAWG Report 2018)
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|}
\hline Survey & Area & Season & 2011 & 2012 & 2013 & 2014 & 2015 & 2016 & 2017 \\
\hline Partial & Lyme Bay* & Oct & 33861 & 24246 & 68613 & 69059 & 10616 & 6120 & 29274 \\
\hline FSP & Lyme Bay** & Oct & 33861 & 27971 & & & & & \\
\hline PELTIC & W Eng Ch & May & 85358 & & & & & & \\
\hline PELTIC & W Eng Ch & Oct & & & 79149 & 93839 & 73574 & 11804 & 34694 \\
\hline
\end{tabular}
* ICES rectangles 29E6, 30E6
-Figure 12.1.1- Figure needs both axis labels. The colors used in the figure (legend) also should be labeled.
-Figure 12.2.1- Figure needs a full-name label in x -axis (e.g. Total length).
-Figure 12.2.2- Please rearrange sub-figures in a time sequence (Nov. in left and Dec. in right) and use full names labels in \(x\) - (e.g. Total length) and \(y\)-axis (e.g. Numbers). The unit of x -axis label should also be added.
-Figure 12.2.3- Figure needs a full-name label in \(x\)-axis (e.g. Total length) and rename the titles of sub-figures (e.g. Quarter 1 and Quarter 4).
-Figure 12.3.2- Figure needs a full-name label in x -axis (e.g. Total length) and a label in y -axis.
-Figure 12.3.1- The figure caption indicated that there were stations designed for zooplankton and oceanographic survey. However, it was not mentioned in the WG report or stock annex. The RG recommends that the information related to zooplankton and oceanographic survey should be added in the WG report.
-Figure 12.6.1- Text is too small to read.
-Figure 12.6.3- Figure needs both axis labels.
-Figure 12.7.1- Figure needs both axis labels.

\section*{10. Conclusions:}
-Given the available data, the assessment of sprat stock in subareas 7.de appears well done. The RG recommends the report be accepted. The major comments are listed below. oThe spatial distribution of the stock was well investigated, but the stock structure for sprat in the English Channel was still unclear. The RG believed that more investigations are required to identify the stock structure for sprat as well as the connection between the Western English Channel stock and Bristol Channel. oIt was unclear whether the zooplankton and oceanographic survey was conducted during the PELTIC acoustic survey. The RG would like to recommend the WG provide related information on that and encourage the WG to explore the relationship between oceanographic condition and large inter-annual fluctuations in stock biomass.
- The RG also recommends the WG apply various analytical assessment methods (e.g. SPiCT with environmental variables) based on best available survey information and develop more appropriate approach for providing catch advice.

\section*{References}

Aldrin, M., Aanes, S., and Subbey, S. 2019. Comments on incongruous formulations in the SAM (state-space assessment model) model and consequences for fish stock assessment. Fisheries Research, 210: 224-227.
Berg, C.W., and Nielsen, A. 2016. Accounting for correlated observations in an age-based state-space stock assessment model. ICES Journal of Marine Science, 73: 1788-1797.
Brooks, E.N., Legault, C.M., and Wilberg, M. 2016. Retrospective forecasting - evaluating performance of stock projections for New England groundfish stocks. Canadian Journal of Fisheries and Aquatic Sciences, 73(6): 935-950.
Butterworth, D.S., and Rademeyer, R.A. 2008. Statistical catch-at-age analysis vs. ADAPTVPA: The case of Gulf of Maine cod. ICES Journal of Marine Science, 65(9): 17171732.

ICES. 2017. Report of the Benchmark Workshop on the Irish Sea Ecosystem (WKIRISH3), 30 January-3 February, Galway, Ireland.
ICES. 2018. ICES Special Request Advice Celtic Seas Ecoregion. EU request to ICES to provide plausible and updated FMSY ranges for the stocks of species inhabiting western EU waters.
Legault, C.M., and Palmer, M.C. 2016. In what direction should the fishing mortality target change when natural mortality increases within an assessment? Canadian Journal of Fisheries and Aquatic Sciences, 73: 349-357.
Marandel, F., Lorance, P., and Trenkei, V.M. 2016. A Bayesian state-space model to estimate population biomass with catch and limited survey data: application to the thornback ray (Raja clavata) in the Bay of Biscay. Aquatic Living Resources, 9(209).
Scott, F., Jardim, E., Millar, C. P., \& Cerviño, S. 2016. An applied framework for incorporating multiple sources of uncertainty in fisheries stock assessments. PloS one, 11(5), e0154922.```


[^0]:    ICES INTERNATIONAL COUNCIL FOR THE EXPLORATION OF THE SEA CIEM CONSEIL INTERNATIONAL POUR L'EXPLORATION DE LA MER

[^1]:    ${ }^{1}$ Special condition: of which up to $10 \%$ may be fished in: 6; Union and international waters of $5 . b$; international waters of 12 and 14 (ANF/*56-14).

[^2]:    $\left({ }^{1}\right)$ By-catch of cod in the area covered by this TAC may be landed provided that it does not comprise more than $1,5 \%$ of the live weight of the total catch retained on board per fishing trip. This provision shall not apply for catches subject to the landing obligation.

[^3]:    ${ }^{1}$ MSS = Marine Scotland Science; MI = Marine Institute Ireland.

[^4]:    * Mesh size 110-119 mm.

[^5]:    * data calculated using estimates from discard observer trips.

[^6]:    * The discards are extrapolated form $30 \%$ of catch in 1985 to $15 \%$ of catch in 2012. Estimates from 2013 onwards are derived from data submitted to InterCatch.

[^7]:    *From 1999 onwards mean weights are shown for trawl and creels combined.

[^8]:    * provisional. **included in UK landings.

[^9]:    * provisional.

[^10]:    ${ }^{1}$ There is a large price differential between the large and small grades. So less volume of the larger grade generates an economically viable return for fishing.

[^11]:    *reduced isometric grid.

[^12]:    * The 2013 survey achieved partial coverage $\sim 60 \%$ of the total area. The abundance has been scaled up to the entire area since densities in the unsurveyed part of the ground were not significantly different in 2014. nr= no reliable abundance estimate could be calculated because survey coverage was partial.

[^13]:    * reduced isometric grid 4.5 nmi .

[^14]:    * Preliminary.

[^15]:    ${ }^{1}$ At the ADGCS 2018, it was agreed to use the SPiCT trends (instead of the survey trends) as an indicator of stock development (see Annex 4 for the review).

[^16]:    * Preliminary landings statistics.

[^17]:    ${ }^{1}$ https://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32018R0120\&from=FR.
    ${ }^{2}$ All types of demersal trawls, including OTB, OTT, PTB, TBB, TBN, TBS, TB.
    ${ }^{3}$ All types of seines, including SSC, SDN, SPR, SV, SB, SX.
    ${ }^{4}$ All longlines or pole and line or rod and line fisheries, including LHP, LHM, LLD, LL, LTL, LX and LLS.
    ${ }^{5}$ All fixed gillnets and traps, including GTR, GNS, FYK, FPN and FIX.

[^18]:    * Average of French and UK release weights.

[^19]:    Landings are preliminary

[^20]:    ${ }^{(1)}$ Exclusively for by-catches of whiting in fisheries for other species. No directed fisheries for whiting are permitted under this quota.

[^21]:    * Estimates for 2019 and 2020 are TSA projections.

[^22]:    * Estimates for 2019 and 2020 are TSA projections.

[^23]:    * Estimates for 2019 and 2020 are TSA projections.

[^24]:    * Estimates for 2019 and 2020 are TSA projections.

[^25]:    ${ }^{(1)}$ Exclusively for by-catches of whiting in fisheries for other species. No directed fisheries for whiting are permitted under this quota.

[^26]:    * Preliminary.

