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# Report of the Working Group on North Atlantic Salmon (WGNAS) 

3-12 April 2013
Copenhagen, Denmark

# International Council for the Exploration of the Sea Conseil International pour l'Exploration de la Mer 

H. C. Andersens Boulevard 44-46<br>DK-1553 Copenhagen V<br>Denmark<br>Telephone (+45) 33386700<br>Telefax (+45) 33934215<br>www.ices.dk<br>info@ices.dk<br>Recommended format for purposes of citation:

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## Executive summary

Working Group on North Atlantic Salmon [WGNAS], ICES HQ, 3-12 April 2013.
Chair: Ian Russell (UK).
Number of participants: 20 representing eleven countries from North America (NAC) and the Northeast Atlantic (NEAC). Information was also provided by correspondence from Greenland, Sweden, Faroes, Denmark, and Spain for use by the Working Group.
WGNAS met to consider questions posed to ICES by the North Atlantic Salmon Conservation Organisation (NASCO). The need for catch advice was dependent on the outcome of applying two indicator frameworks prior to the meeting.

- In 2012, the Working Group advised that there were no mixed-stock fishery options at West Greenland in 2012 to 2014 nor in NAC in 2012 to 2105 that would be consistent with a $75 \%$ chance or greater of simultaneously meeting the seven (for West Greenland) and six (for NAC) management objectives for 2SW salmon. The West Greenland Framework of indicators was applied in January 2013 and did not indicate the need for an updated assessment of catch options and no new management advice for this fishery was requested by NASCO.
- A Framework of Indicators (FWI) was developed for NEAC stocks in 2012 and was also applied in January 2013 in relation to the multi-annual agreement for the Faroes fishery. This indicated that the forecasted pre fishery abundance (PFA) for one of the stock complexes (Southern NEAC MSW fish) may have been overestimated; this triggered a request from NASCO for a reassessment of the stocks and an update to the catch advice.

The terms of reference were addressed by reviewing working documents prepared ahead of the meeting as well as the development of documents and text for the report during the meeting. The report is structured by sections specific to the terms of reference of the WGNAS.

- In the North Atlantic, exploitation rates have declined and nominal catch of wild Atlantic salmon in 2012 was 1409 t , the second lowest in the time-series beginning in 1960.
- The Working Group reported on a range of new opportunities for salmon assessment and management (e.g. modelling developments, fish tracking technologies, genetic investigations) and potential threats (e.g. parasites, artificial light). The Working Group reviewed the potential threat to Atlantic salmon posed by exotic salmonid species.
- The four NEAC stock complexes had a greater than 95\% probability of having exceeded their conservation limits (CLs) in 2012 and were therefore considered to be at full reproductive capacity prior to the commencement of distant water fisheries. However at a country level, stocks from several jurisdictions were below CLs.
- The risk based framework for the provision of catch advice for the Faroes Fishery developed in 2012 at the NEAC stock complex level was run at both stock complex and country level.
- There are no catch options for the Faroes fishery that would allow all national or stock complex management units to achieve their CLs with a greater than $95 \%$ probability in any of the seasons 2013/2014 to 2015/2016.
- The NEAC FWI was updated and the Working Group recommends that a slight change is made to its future operation; such that a one-tailed approach is used where the fishery is closed (i.e. no reassessment is signalled where the FWI suggests a further reduction in abundance). This would have avoided the need for a reassessment in 2013.
- North American 2SW spawner estimates were below their CLs in each of the six regions. Within each of the geographic areas there are also varying numbers of individual river stocks which are failing to meet CLs, particularly in the southern areas of Scotia-Fundy and the USA. In 2012, large declines in abundance were noted from the higher abundances noted in 2011 reflecting increased mortality at sea on 1SW and 2SW salmon.
- There was a catch of $33 t$ in the fishery at Greenland in 2013. The overall abundance of salmon within the West Greenland area remains low relative to historical levels and six of the seven stock complexes exploited in the fishery are below CLs.
- Marine survival indices in the North Atlantic have improved in some index stocks in recent years, but the declining trend has persisted and survival indices remain low. Factors other than marine fisheries, acting in freshwater and in the ocean in both NAC and NEAC areas (e.g. marine mortality, fish passage, water quality) are contributing to continued low abundance of wild Atlantic salmon.


## 1 Introduction

### 1.1 Main tasks

At its 2012 Statutory Meeting, ICES resolved (C. Res. 2012/2/ACOM09) that the Working Group on North Atlantic Salmon [WGNAS] (chaired by: Ian Russell, UK) will meet at ICES HQ, 3-12 April 2013 to consider questions posed to ICES by the North Atlantic Salmon Conservation Organization (NASCO).

The terms of reference were met and the sections of the report which provide the answers are identified below:

| a) With respect to Atlantic Salmon in the North Atlantic area: | Section 2 |
| :---: | :---: |
| i) provide an overview of salmon catches and landings, including unreported catches by country, catch and release, and production of farmed and ranched Atlantic salmon in 2012 ${ }^{1}$; | 2.1 and 2.2 <br> Annex 4 |
| ii) report on significant new or emerging threats to, or opportunities for, salmon conservation and management ${ }^{2}$; | 2.3 |
| iii) provide a review of examples of successes and failures in wild salmon restoration and rehabilitation and develop a classification of activities which could be recommended under various conditions or threats to the persistence of populations; | 2.4 |
| iv) advise on the potential threats to Atlantic salmon from exotic salmonids including brown trout and rainbow trout where appropriate; | 2.5 |
| v) provide a compilation of tag releases by country in 2012; | 2.6 |
| vi) identify relevant data deficiencies, monitoring needs and research requirements. Where relevant suggest improvement for the revision of the DCF, to be taken into account by WKESDCF. | $2.7$ <br> Annex 8 |


| b) With respect to Atlantic salmon in the Northeast Atlantic Commission area: | Section 3 |
| :--- | :--- |
| i) describe the key events of the 2012 fisheries ${ }^{3}$; | 3.1 |
| ii) review and report on the development of age-specific stock conservation limits; | 3.2 |
| iii) describe the status of the stocks; | 3.3 |
| iv) further develop a risk-based framework for the provision of catch advice for the Faroese <br> salmon fishery reporting on the implications of selecting different numbers of management <br> units ${ }^{4} ;$ | 3.4 |
| In the event that NASCO informs ICES that the Framework of Indicators (FWI) indicates that <br> reassessment is required: * |  |
| v) provide catch options or alternative management advice for 2013-2016, with an <br> assessment of risks relative to the objective of exceeding stock conservation limits and <br> advise on the implications of these options for stock rebuilding; | $3.5,3.6$ |
| vi) update the Framework of Indicators used to identify any significant change in the <br> previously provided multi-annual management advice. | 3.7 |


| c) With respect to Atlantic salmon in the North American Commission area: | Section 4 |
| :--- | :--- |
| i) describe the key events of the 2012 fisheries (including the fishery at St Pierre and <br> Miquelon) |  |
| ii) update age-specific stock conservation limits based on new information as available; | 4.1 |

iii) describe the status of the stocks;
In the event that NASCO informs ICES that the Framework of Indicators (FWI) indicates that
reassessment is required: *
reasessment is required:*
iv) provide catch options or alternative management advice for 2013-2016 with an assessment of risks relative to the objective of exceeding stock conservation limits and advise on the implications of these options for stock rebuilding;
v) update the Framework of Indicators used to identify any significant change in the previously provided multi-annual management advice.

| d) With respect to Atlantic salmon in the West Greenland Commission area: | Section 5 |
| :--- | :--- |
| i) describe the key events of the 2012 fisheries $^{3} ;$ | 5.1 |
| ii) Describe the status of the stocks ${ }^{6} ;$ | 5.2 |

In the event that NASCO informs ICES that the Framework of Indicators (FWI) indicates that reassessment is required: *
iii) provide catch options or alternative management advice for 2013-2015 with an assessment of risk relative to the objective of exceeding stock conservation limits and advise on the implications of these options for stock rebuilding ${ }^{5}$;
iv) update the Framework of Indicators used to identify any significant change in the previously provided multi-annual management advice.

1. With regard to question a) $\mathbf{i}$, for the estimates of unreported catch the information provided should, where possible, indicate the location of the unreported catch in the following categories: in-river; estuarine; and coastal. Numbers of salmon caught and released in recreational fisheries should be provided.
2. With regard to question a) $\mathbf{i i}$, ICES is requested to include reports on any significant advances in understanding of the biology of Atlantic salmon that is pertinent to NASCO, including information on any new research into the migration and distribution of salmon at sea and the potential implications of climate change for salmon management.
3. In the responses to questions b) i, c) i and d) i, ICES is asked to provide details of catch, gear, effort, composition and origin of the catch and rates of exploitation. For homewater fisheries, the information provided should indicate the location of the catch in the following categories: in-river; estuarine; and coastal. Any new information on non-catch fishing mortality of the salmon gear used, on the bycatch of other species in salmon gear, and on the bycatch of salmon in any existing and new fisheries for other species is also requested.
4. In response to question b) iv, ICES is asked to advise on the limitations for defining management units smaller than the current NEAC stock complexes, the implications of applying probabilities of achieving CLs to separate management units vs. the use of simultaneous probabilities and the choice of risk levels for achieving management objectives.
5. In response to questions b) v, c) iv and d) iii, provide a detailed explanation and critical examination of any changes to the models used to provide catch advice and report on any developments in relation to incorporating environmental variables in these models.
6. In response to question d) ii, ICES is requested to provide a brief summary of the status of North American and Northeast Atlantic salmon stocks. The detailed information on the status of these stocks should be provided in response to questions b) iii and c) iii.

* The aim should be for NASCO to inform ICES by 31 January of the outcome of utilizing the FWI.

The NEAC FWI indicated that a reassessment was required (on the grounds that PFA had been overestimated for one stock complex), while the West Greenland FWI signalled no reassessment was necessary. Questions b) v and vi were therefore addressed by the Working Group during the 2013 meeting; there was no requirement to address questions c) iv and $v$ or d) iii and iv.

In response to the Terms of Reference, the Working Group considered 38 Working Documents submitted by participants (Annex 1); other references cited in the Report are given in Annex 2. Additional information was supplied by Working Group members unable to attend the meeting by correspondence. A full address list for the meeting participants is provided in Annex 3. A complete list of acronyms used within this document is provided in Annex 7.

### 1.2 Participants

| Member | Country |
| :--- | :--- |
| Chaput, G. | Canada |
| Ensing, D. | UK (N. Ireland) |
| Erkinaro, J. | Finland |
| Euzenat, G. | France |
| Fiske, P. | Norway |
| Gjøsæeter, H. | Norway |
| Gudbergsson, G. | Iceland |
| Massiot-Granier, F. | France |
| Meerburg, D. | Canada |
| Nygaard, R. | Ireland |
| Ó Maoiléidigh, N. | UK (Scotland) (by WebEx) |
| Orpwood, J. | UK (England \& Wales) |
| Potter, T. | Russia |
| Prusov, S. | France |
| Rivot, E. | Canada |
| Robertson, M. | UK (England \& Wales) |
| Russell, I. (Chair) | USA (Scotland) |
| Sheehan, T. | Ireland |
| Smith, G. W. | Ustyuzhinskiy, G. |

### 1.3 Management framework for salmon in the North Atlantic

The advice generated by ICES is in response to Terms of Reference posed by the North Atlantic Salmon Conservation Organization (NASCO), pursuant to its role in international management of salmon. NASCO was set up in 1984 by international convention (the Convention for the Conservation of Salmon in the North Atlantic Ocean), with a responsibility for the conservation, restoration, enhancement, and rational
management of wild salmon in the North Atlantic. While sovereign states retain their role in the regulation of salmon fisheries for salmon originating in their own rivers, distant water salmon fisheries, such as those at Greenland and Faroes, which take salmon originating in rivers of another Party are regulated by NASCO under the terms of the Convention. NASCO now has six Parties that are signatories to the Convention, including the EU which represents its Member States.

NASCO discharges these responsibilities via three Commission areas shown below:


### 1.4 Management objectives

NASCO has identified the primary management objective of that organization as:
"To contribute through consultation and cooperation to the conservation, restoration, enhancement and rational management of salmon stocks taking into account the best scientific advice available".

NASCO further stated that "the Agreement on the Adoption of a Precautionary Approach states that an objective for the management of salmon fisheries is to provide the diversity and abundance of salmon stocks" and NASCOs Standing Committee on the Precautionary Approach interpreted this as being "to maintain both the productive capacity and diversity of salmon stocks" (NASCO, 1998).

NASCO's Action Plan for Application of the Precautionary Approach (NASCO, 1999) provides interpretation of how this is to be achieved, as follows:

- "Management measures should be aimed at maintaining all stocks above their conservation limits by the use of management targets".
- "Socio-economic factors could be taken into account in applying the Precautionary Approach to fisheries management issues".
- "The precautionary approach is an integrated approach that requires, inter alia, that stock rebuilding programmes (including, as appropriate, habitat improvements, stock enhancement, and fishery management actions) be developed for stocks that are below conservation limits".


### 1.5 Reference points and application of precaution

Conservation limits (CLs) for North Atlantic salmon stock complexes have been defined by ICES as the level of stock (number of spawners) that will achieve long-term average maximum sustainable yield (MSY). In many regions of North America, the CLs are calculated as the number of spawners required to fully seed the wetted area of the river. In some regions of Europe, pseudo stock-recruitment observations are used to calculate a hockey-stick relationship, with the inflection point defining the CLs. In the remaining regions, the CLs are calculated as the number of spawners that will achieve long-term average maximum sustainable yield (MSY), as derived from the adult-to-adult stock and recruitment relationship (Ricker, 1975; ICES, 1993). NASCO has adopted the region specific CLs (NASCO 1998). These CLs are limit reference points (Slim); having populations fall below these limits should be avoided with high probability.

Atlantic salmon has characteristics of short-lived fish stocks; mature abundance is sensitive to annual recruitment because there are only a few age groups in the adult spawning stock. Incoming recruitment is often the main component of the fishable stock. For such fish stocks, the ICES MSY approach is aimed at achieving a target escapement (MSY Bescapement, the amount of biomass left to spawn). No catch should be allowed unless this escapement can be achieved. The escapement level should be set so there is a low risk of future recruitment being impaired, similar to the basis for estimating $B_{p a}$ in the precautionary approach. In short-lived stocks, where most of the annual surplus production is from recruitment (not growth), MSY Bescapement and $B_{p a}$ might be expected to be similar.

It should be noted that this is equivalent to the ICES precautionary target reference points ( $\mathrm{S}_{\mathrm{pa}}$ ). Therefore, stocks are regarded by ICES as being at full reproductive capacity only if they are above the precautionary target reference point. This approach parallels the use of precautionary reference points used for the provision of catch advice for other fish stocks in the ICES area.

Management targets have not yet been defined for all North Atlantic salmon stocks. When these have been defined they will play an important role in ICES advice.

For the assessment of the status of stocks and advice on management of national components and geographical groupings of the stock complexes in the NEAC area, where there are no specific management objectives:

- ICES requires that the lower bound of the confidence interval of the current estimate of spawners is above the CL for the stock to be considered at full reproductive capacity.
- When the lower bound of the confidence limit is below the CL, but the midpoint is above, then ICES considers the stock to be at risk of suffering reduced reproductive capacity.
- Finally, when the midpoint is below the CL, ICES considers the stock to suffer reduced reproductive capacity.

For catch advice on fish exploited at West Greenland (non-maturing 1SW fish from North America and non-maturing 1SW fish from Southern NEAC), ICES has adopted, a risk level of $75 \%$ of simultaneous attainment of management objectives (ICES, 2003) as part of an agreed management plan. ICES applies the same level of risk aversion for catch advice for homewater fisheries on the North American stock complex.

## 2 Atlantic salmon in the North Atlantic area

### 2.1 Catches of North Atlantic salmon

### 2.1.1 Nominal catches of salmon

The nominal catch of a fishery is defined as the round, fresh weight of fish that are caught and retained. Total nominal catches of salmon reported by country in all fisheries for 1960-2012 are given in Table 2.1.1.1. Catch statistics in the North Atlantic also include fish-farm escapees and, in some Northeast Atlantic countries, ranched fish (see Section 2.2.2). Catch and release has become increasingly commonplace in some countries, but these fish do not appear in the nominal catches (see Section 2.1.2).

Icelandic catches have traditionally been split into two separate categories, wild and ranched, reflecting the fact that Iceland has been the only North Atlantic country where large-scale ranching has been undertaken with the specific intention of harvesting all returns at the release site. The release of smolts for commercial ranching purposes ceased in Iceland in 1998, but ranching for rod fisheries in two Icelandic rivers continued into 2012 (Table 2.1.1.1). While ranching does occur in some other countries, this is on a much smaller scale. Some of these operations are experimental and at others harvesting does not occur solely at the release site. The ranched component in these countries has therefore been included in the nominal catch.

Figure 2.1.1.1 shows the total reported nominal catch of salmon grouped by the following areas: ‘Northern Europe’ (Norway, Russia, Finland, Iceland, Sweden and Denmark); 'Southern Europe' (Ireland, UK (Scotland), UK (England \& Wales), UK (N. Ireland), France and Spain); 'North America' (Canada, USA and St Pierre et Miquelon (France)); and 'Greenland and Faroes'.

The provisional total nominal catch for 2012 was $1409 \mathrm{t}, 220 \mathrm{t}$ below the updated catch for 2011 (1629 t). The 2012 catch was below the average of the previous five years (1565 t) and below the average of the last ten years (1928 t). Catches were below the previous tenyear averages in the majority of Southern NEAC and Northern NEAC countries, except Sweden and Finland.

Nominal catches in homewater fisheries were split, where available, by sea age or size category (Table 2.1.1.2 weight only). The data for 2012 are provisional and, as in Table 2.1.1.1, include both wild and reared salmon and fish-farm escapees in some countries. A more detailed breakdown, providing both numbers and weight for different sea age groups for most countries, is provided at Annex 4. Countries use different methods to partition their catches by sea age class (outlined in the footnotes to Annex 4). The composition of catches in different areas is discussed in more detail in Sections 3, 4, and 5.

ICES recognizes that mixed-stock fisheries present particular threats to stock status. These fisheries predominantly operate in coastal areas and NASCO specifically requests that the nominal catches in homewater fisheries be partitioned according to whether the catch is taken in coastal, estuarine or riverine areas. Figure 2.1.1.2 presents these data on a country-by-country basis. It should be noted, however, that the way in which the nominal catch is partitioned among categories varies between countries, particularly for estuarine and coastal fisheries. For example, in some countries these catches are split according
to particular gear types and in other countries the split is based on whether fisheries operate inside or outside headlands. While it is generally easier to allocate the freshwater (riverine) component of the catch, it should also be noted that catch and release is now in widespread use in several countries (Section 2.1.2) and these fish are excluded from the nominal catch. Noting these caveats, these data are considered to provide the best available indication of catch in these different fishery areas. Figure 2.1.1.2 shows that there is considerable variability in the distribution of the catch among individual countries. There are no coastal fisheries in Iceland, Finland and coastal fisheries ceased in Ireland in 2007. The coastal catch has declined markedly in UK (N. Ireland). In most countries the majority of the catch is now taken in freshwater except UK (England \& Wales), Norway and Russia where roughly half of the total catch is still taken in coastal waters.

Coastal, estuarine and riverine catch data for the period 2002 to 2012 aggregated by region are presented in Figure 2.1.1.3. In northern Europe, catches in coastal fisheries have been in decline over the period and reduced from 663 t in 2002 to 300 t in 2012. Freshwater catches have been relatively constant. At the beginning of the time-series about half the catch was taken in rivers and half in coastal waters. The proportion of the catch taken in coastal waters over the last five years represents only one third of the total. In southern Europe, catches in all fishery areas have declined dramatically over the period. While coastal fisheries have historically made up the largest component of the catch, these fisheries have declined the most from 718 t in 2002 to 68 t in 2012, reflecting widespread measures to reduce exploitation in a number of countries. In the last six years, the majority of the catch in this area has been taken in freshwater, though there was a slight increase in the proportion of the catch taken in coastal waters in 2010 and 2011.

In North America, the total catch has been fluctuating over the period 2002 to 2012. The majority of the catch in this area has been taken in riverine fisheries; the catch in coastal fisheries has been relatively small in any year with the biggest catch taken in $2011(14 \mathrm{t})$.

### 2.1.2 Catch and release

The practice of catch and release in rod fisheries has become increasingly common as a salmon management/conservation measure in light of the widespread decline in salmon abundance in the North Atlantic. In some areas of Canada and USA, catch and release has been practiced since 1984, and in more recent years it has also been widely used in many European countries both as a result of statutory regulation and through voluntary practice.

The nominal catches presented in Section 2.1.1 do not include salmon that have been caught and released. Table 2.1.2.1 presents catch-and-release information from 1991 to 2012 for countries that have records. Catch and release may also be practiced in other countries while not being formally recorded. There are large differences in the percentage of the total rod catch that is released: in 2012 this ranged from 14\% in Norway (this is a minimum figure) to $74 \%$ in UK (Scotland) reflecting varying management practices and angler attitudes among these countries. Catch and release rates have typically been highest in Russia (average of $84 \%$ in the five years 2004 to 2008) and are believed to have remained at this level. However, there were no obligations to report caught-and-released fish in Russia since 2009. Within countries, the percentage of fish released has tended to increase over time; however there was a slight decrease in numbers reported in some countries in 2012. There is also evidence from some countries that larger MSW fish are
released in larger proportions than smaller fish. Overall, over 173000 salmon were reported to have been released around the North Atlantic in 2012, slightly below the average of the last five years (188 000).

Summary information on how catch and release levels are incorporated into national assessments was provided to the Working Group in 2010 (ICES, 2010b).

### 2.1.3 Unreported catches

Unreported catches by year (1987 to 2012) and Commission Area are presented in Table 2.1.3.1 and are presented relative to the total nominal catch in Figure 2.1.3.1. A description of the methods used to derive the unreported catches was provided in ICES (2000) and updated for the NEAC Region in ICES (2002). Detailed reports from different countries were also submitted to NASCO in 2007 in support of a special session on this issue. There have been no estimates of unreported catch for Russia since 2008 and for Canada in 2007 and 2008. Estimates for Canada in 2009 and 2010 were considered incomplete (information available for three of the four jurisdictions). There are also no estimates of unreported catch for Spain and St Pierre et Miquelon (NAC), where total catches are typically small. It has not typically been possible to separate the unreported catch into that taken in coastal, estuarine and riverine areas.

In general, the derivation methods used by each country have remained relatively unchanged and thus comparisons over time may be appropriate. However, the estimation procedures vary markedly between countries. For example, some countries include only illegally caught fish in the unreported catch, while other countries include estimates of unreported catch by legal gear as well as illegal catches in their estimates. Over recent years efforts have been made to reduce the level of unreported catch in a number of countries (e.g. through improved reporting procedures and the introduction of carcass tagging and logbook schemes).

The total unreported catch in NASCO areas in 2012 was estimated to be 403 t . The unreported catch in the Northeast Atlantic Commission Area in 2012 was estimated at 363 t , and that for the West Greenland and North American Commission Areas at 10 t and 31 t , respectively. The 2012 unreported catch by country is provided in Table 2.1.3.2. Information on unreported catches was not provided to enable these to be partitioned into coastal, estuarine and riverine areas.

In the past, salmon fishing by non-contracting parties is known to have taken place in international waters to the north of the Faroe Islands. Typically, a number of surveillance flights have taken place over this area in recent years. In 2012, there were 15 surveillance flights by the Norwegian coastguard over the area of international waters in the Norwegian Sea between the beginning of April and end of October. As in past years there were no sightings of vessels fishing for salmon, although there have been extended periods over the winter period when no flights took place. This is a period when salmon fishing has previously been reported.

Summary information on how unreported catches are incorporated into national and international assessments was provided to the Working Group in 2010 (ICES, 2010b).

### 2.2 Farming and sea ranching of Atlantic salmon

### 2.2.1 Production of farmed Atlantic salmon

The provisional estimate of farmed Atlantic salmon production in the North Atlantic area for 2012 is 1450 kt . The production of farmed Atlantic salmon in this area has been over one million tonnes since 2009. The 2012 total represents an $8 \%$ increase on 2011 and a $30 \%$ increase on the previous five-year mean (Table 2.2.1.1 and Figure 2.2.1.1) due to increased production in the majority of countries where farming occurs. Norway and UK (Scotland) continue to produce the majority of the farmed salmon in the North Atlantic (79\% and $11 \%$ respectively). Farmed salmon production in 2012 was above the previous fiveyear average in all countries. Data for UK (N. Ireland) since 2001 and data for USA since 2011 are not publicly available.

Worldwide production of farmed Atlantic salmon has been over one million tonnes since 2002 and approached two million tonnes in 2012. It is difficult to source reliable production figures for all countries outside the North Atlantic area and it has been necessary to use 2011 estimates for some countries in deriving a worldwide estimate for 2012. Noting this caveat, total production in 2012 is provisionally estimated at around 1961 kt (Table 2.2.1.1 and Figure 2.2.1.1), a $6 \%$ increase on 2011. Production of farmed Atlantic salmon outside the North Atlantic is estimated to have accounted for $26 \%$ of the total in 2012 (up from $20 \%$ in 2011). Production outside the North Atlantic is still dominated by Chile and is now in excess of what it was prior to the outbreak of infectious salmon anaemia (ISA) which impacted the industry in that country from 2007.

The worldwide production of farmed Atlantic salmon in 2012 was over 1300 times the reported nominal catch of Atlantic salmon in the North Atlantic.

### 2.2.2 Harvest of ranched Atlantic salmon

Ranching has been defined as the production of salmon through smolt releases with the intent of harvesting the total population that returns to freshwater (harvesting can include fish collected for broodstock) (ICES, 1994). The release of smolts for commercial ranching purposes ceased in Iceland in 1998, but ranching with the specific intention of harvesting by rod fisheries has been practiced in two Icelandic rivers since 1990 and these data have now been included in the ranched catch (Table 2.1.1.1). The total harvest of ranched Atlantic salmon in countries bordering the North Atlantic in 2012 was 12 t, all taken by the Icelandic ranched rod fisheries (Table 2.2.2.1; Figure 2.2.2.1). Small catches of ranched fish from experimental projects were also known for Ireland but data were not available for 2012. No estimate of ranched salmon production was made in Norway in 2012 where such catches have been very low in recent years ( $<1 \mathrm{t}$ ) and UK (N. Ireland) where the proportion of ranched fish was not assessed between 2008 and 2012 due to a lack of microtag returns.

It was noted that a large proportion of the fish caught in Sweden in the last ten years ( $66 \%$ of the total catch in 2012 ( 20 t ) and $69 \%$ of the previous ten-year average) originated from hatchery-reared smolts released under programmes to mitigate for hydropower development schemes. However, these fish do not fall within the agreed definition of ranched fish and are not included in Figure 2.2.2.1.

### 2.3 NASCO has asked ICES to report on significant, new or emerging threats to, or opportunities for, salmon conservation and management

### 2.3.1 Dam Impact Analysis Model for Atlantic Salmon in the Penobscot River, Maine

The Dam Impact Analysis (DIA) model is a population viability analysis that was developed to help better understand the impacts of dams on the production potential of Atlantic salmon (Nieland et al., 2013). Dams have been identified as a major contributor to the historic decline and current low abundance of salmon in the Gulf of Maine Distinct Population Segment, which was first listed as endangered in 2000 and then expanded in 2009. The DIA model specifically simulates the interactions of Atlantic salmon and 15 hydroelectric dams in the Penobscot River watershed in Maine, USA (Figure 2.3.1.1).
A life-history modelling approach was undertaken to incorporate life stage-specific information for Atlantic salmon. DIA model inputs (obtained from either field data or literature reports) were used to simulate the life cycle of Atlantic salmon in the Penobscot River (Figure 2.3.1.2). Most model inputs were year- and iteration-specific random draws from distributions to incorporate stochastic variation into the model, and Monte Carlo sampling was used to simulate many iterations of this population forward in time. Several modelling scenarios were run to reflect recent conditions in the Penobscot River (i.e. prior to the planned removal of specific dams) as well as possible future conditions. Adult abundance, distribution of adults throughout the watershed, and number and proportion of smolts killed by dam-induced mortality were used as performance metrics for each scenario.

The modelled population of Atlantic salmon in the Penobscot River decreased in abundance and distribution when DIA model inputs were set to reflect recent conditions (i.e. Base Case), whereas abundance increased and Atlantic salmon remained distributed throughout the Penobscot River watershed when marine and freshwater survival rates were increased by four and by two times recent levels (i.e. Recovery), respectively (Figure 2.3.1.3). The production potential of Atlantic salmon was also affected by mainstem dams more than tributary dams in the Penobscot River watershed (Figure 2.3.1.3). Sensitivity analyses were performed on all input values to determine which model inputs had the greatest impact on the results. The DIA model results were most sensitive to the marine survival and downstream dam passage survival rates.

The DIA model can project changes in future abundance and can provide information about model inputs that can help inform recovery efforts for the modelled population. The model is not meant to predict absolute abundance, distribution, or mortality but should instead be used to evaluate the relative changes in the Penobscot River population of Atlantic salmon under different modelling scenarios.

### 2.3.2 Marine influences on North American Atlantic salmon populations

The population abundance and marine survival rates of Atlantic salmon have declined throughout their range, and limitations in our understanding of factors responsible for these declines have been widely recognized (Hansen et al., 2012). A new study investigating Atlantic salmon population declines across North America and how these declines
have been shaped by marine ecosystem conditions has recently been completed (Mills et al., 2013).

While patterns of declines in abundance and productivity appear similar (Figure 2.3.2.1), this study confirmed through dynamic factor analysis that abundance and productivity of Atlantic salmon populations changed in a coherent manner across major regions of North America from the US to Labrador. Populations showing coherent trends are likely to be responding to common factors, which North American Atlantic salmon would experience during the marine portion of their life. Thus, the observed coherence among populations points towards a likely shift in marine survival and strong influence of marine conditions.

Major changes in Atlantic salmon population characteristics were detected after 1990 and 1997 based on a chronological cluster analysis (Figure 2.3.2.2), and these population shifts could be linked to changes in climate, physical, and biological conditions in the marine ecosystem. The decline in salmon abundance after 1990 was preceded by a series of changes across multiple levels of the ecosystem, including climate indices (i.e. AMO and NAO); physical conditions such as temperature and salinity; and biological characteristics such as phytoplankton abundance, zooplankton community composition, and capelin size. A subsequent shift in salmon productivity in 1997 followed an unusually low NAO event in 1996.

Pairwise associations between Atlantic salmon population trends and a suite of climate, physical, and biological factors were further investigated to understand how marine ecosystem changes may be related to Atlantic salmon. Results of these analyses indicate that climate conditions can be directly linked to the abundance and productivity of North American Atlantic salmon populations; however, many climate and physical influences also act through lower biological trophic levels and thereby indirectly affect Atlantic salmon (Figure 2.3.2.3). The strongest correlations were between salmon and capelin, sea surface temperature, and zooplankton. These results suggest that poor trophic conditions and warming water temperatures throughout their marine habitat area are constraining productivity and recovery of North American Atlantic salmon populations.

### 2.3.3 West Greenland foraging ecology and implications for survival

Declining Atlantic salmon populations throughout the species range, despite diverse population structures and management regimes, suggests reduced productivity and survival during the marine phase. During this phase, fish from North America and Europe congregate at common feeding grounds (e.g. the Norwegian Sea, the Labrador Sea, West Greenland, etc.) to consume abundant energy rich prey that promote rapid growth and sexual maturity. Stomach samples were collected from a total of 1345 salmon from 20062007 and from 2009-2011 (as part of SALSEA West Greenland) from various communities along the west coast of Greenland. While annual variations in the stomach content weights and composition were documented, Atlantic salmon consumed primarily capelin and Themisto sp. (amphipod); a finding consistent with historic data collected from the offshore waters of West Greenland during the 1960s (Figure 2.3.3.1).

Energy equivalents ( $\mathrm{kJ} / \mathrm{g}$ wet weight) of the stomach contents were obtained from values in the literature (Thayer et al., 1973; Steimle and Terranova, 1985; Lawson et al., 1998; Hislop et al., 1991) and were applied to the stomach contents of each individual fish to
estimate the total energy content in a given stomach. These estimates were scaled to body weight for each fish $\left(\mathrm{kJ} \bullet \mathrm{g}^{-1}\right)$ at the time of sampling. The standardized energy content varied annually (Figure 2.3.3.2; ANOVA; $\mathrm{F}_{[3,887]}=42.92, \mathrm{p}<0.001$ ). All post hoc pairwise comparison differences were highly significant between all pairs except differences between 2006 and $2010(\mathrm{p}=0.079)$.

Since capelin is an important resource for Atlantic salmon, and the quality of this resource is decreasing (Figure 2.3.3.3), effects on potential 2SW spawners from southern populations that feed on capelin at Greenland may be exacerbated. These data provide insights into the current foraging conditions off West Greenland. They also complement the documentation of the sharp stock productivity reduction that began in the early 1990s, and provide an opportunity to explore various hypotheses to examine the role energy resources play in the viability of various life-history strategies in the marine phase.

### 2.3.4 Tracking and acoustic tagging studies in Greenland and Canada

### 2.3.4.1 Tagging adult Atlantic salmon at West Greenland with pop-up archival satellite tags (PSATs)

Return rates of the 2SW component of Atlantic salmon populations are decreasing, especially at southern latitudes and on both sides of the North Atlantic. These fish are present at West Greenland, foraging during summer/autumn, prior to initiating a return migration to homewaters to spawn. To investigate the migration dynamics of this population, pop-up archival satellite tags (PSATs) were attached to 25 Atlantic salmon near Nuuk, West Greenland in September 2010-2012, to track them over autumn and winter. Preliminary results suggest that two tags remained on the fish until the programmed pop-off date (April 1), three fish were predated, one tag popped-off due to exceeding the emergency pop-off depth, seven popped-off for unknown reasons and 12 did not transmit any data.

Detailed information on migration routes, migration rates, environmental conditions experienced and habitat preferences can be obtained from the gathered data. The one full term fish from 2010 moved north to Disko Bay immediately after tagging and its swimming depth was constrained while it remained over the shelf waters (Figure 2.3.4.1). After migrating into the Labrador Sea in February (as the sea ice progressed south or after it entered sub $-0^{\circ} \mathrm{C}$ water) it began to dive to depths exceeding 750 m , possibly in search of food.

Three incidences of likely predation were also documented based on inferences from the data received and the capabilities of the tags themselves. It is hypothesized that a Greenland shark, a large Atlantic halibut and an unknown predator consumed three individual salmon (Figure 2.3.4.2). Evidence of predation was determined based on significant changes in the dive profiles, recorded temperatures and the absence of any light intensity records (i.e. the tags did not detect any light indicating they were in the stomach of an animal).

While fish generally used a broad area of the Labrador Sea, overall movement after tagging indicate both northerly and southerly movement patterns along the coast, possibly in search of capelin or other prey resources. Overall, it appears that salmon movements,
behaviour, and thermal habitat requirements are either very individualistic or Atlantic salmon are capable of tolerating regimes well outside those believed optimal. Eight incidences of PSATs popping-off before the programmed release date were documented. These may have been the result of tag/bracket failure or predation events resulting from tag dislodgement given the relatively short duration they were on the fish.

PSAT technology is generally suitable for work with Atlantic salmon of the size range at West Greenland. Atlantic salmon were successfully captured, tagged, released and tracked over winter. High quality data on movement patterns, migration behaviours, winter locations, and conditions experienced at winter habitats were collected. Refining the catch methods to reduce stress, tagging techniques, and determining reasons for failures appears to be critical for long-term success of tagging based on some of these preliminary findings and the given cost of the equipment. These data are not obtainable by other means and may provide valuable information related to a critical understudied life stage of the species, aiding in the conservation and management of Atlantic salmon across the Northwest Atlantic.

### 2.3.4.2 Acoustic tracking update for Canada

The Working Group reviewed the results of ongoing projects, led by the Atlantic Salmon Federation (ASF), to assess estuarine and marine survival of tagged Atlantic salmon released in rivers of the Gulf of St Lawrence. A total of 291 smolts from four rivers in Canada (42 St Jean, 64 Cascapedia, 105 Restigouche, 80 Miramichi) and 35 Miramichi kelts were sonically tagged between April and June 2012. Of the 35 kelts tagged with acoustic tags, ten were tagged with archival pop-up tags. These archival pop-up tags were set to release after four months.

The proportion of smolts detected (apparent survival) in 2012 from freshwater release points to the heads of tide, through the estuary and out of the Strait of Belle Isle was similar to previous years for the Cascapedia, Restigouche and Miramichi rivers; few St Jean fish were detected as previous years. As previously, smolts and kelts exited the Strait of Belle Isle together, however in 2012, the timing occurred approximately two weeks earlier (the last week of June and first week of July). Analysis is proceeding using modelling methodology to account for the variability in detection by receivers so stage survival estimates and their variability may be estimated (see Section 2.3.4.3).

The detector array across the Cabot Strait, between Cape Breton, Nova Scotia and Southwest Newfoundland was completed by the OCEAN Tracking Network (OTN) and functional in 2012 although few fish used this exit from the Gulf of St Lawrence (one Miramchi kelt in late May and one Miramichi smolt in mid-June). The satellite archival pop-up tags functioned well in 2012, with information from seven of the nine that left the Miramichi River being recovered. This information is still being analysed however preliminary results show evidence of predation on a kelt within the Gulf of St Lawrence (likely by a species such as a porbeagle shark) and one fish leaving the Gulf of St Lawrence through the Strait of Belle Isle. The remainder stayed within the Gulf, possibly as a result of predation, or they could be returning to spawn in the same year as they were tagged as consecutive repeat spawners. These salmon may not migrate far from their spawning river.

For the first time in 2012, new modes of detection of acoustically tagged salmon were investigated in the Gulf of St Lawrence in collaboration with the Ocean Tracking Network and DFO. Bioprobe receivers have been mounted on grey seals by DFO; these mammals return annually to Sable Island and at least one seal receiver had detected at least two of ASF's tagged salmon within the Gulf of St Lawrence. The Ocean Tracking Network released a Wave Glider into the Gulf of St Lawrence along the west coast of Newfoundland in late June and the movements of the Wave Glider were controlled to pass through areas expected to contain acoustically tagged smolts and kelts on their migration through the Strait of Belle Isle. Detection of at least one of these fish did occur, however it is believed that the Wave Glider path may not have coincided with the migration as it seemed to occur about two weeks earlier as mentioned earlier.

As detailed in Section 2.8.1 relating to new marine research initiatives, the Working Group also encourages the continuation of this tracking programme as information from it is expected to be useful in the assessment of marine mortality on North American salmon stocks.

### 2.3.4.3 Modelling inter-stage survival rates and detection probabilities for acoustically tracked Atlantic salmon smolts and post-smolts: model, assumptions, diagnostics, considerations for planning experiments

In Section 2.8, the Working Group was informed of the NASCO IASRB Subgroup report on Future Direction of Research on Marine Survival of Salmon. The subgroup indicated that consideration of studies to partition marine mortality of salmon among the phases of the marine migration remained a priority. An outline proposal for using acoustic technologies to tag emigrating smolts from NAC and NEAC as well as adults at West Greenland and to track their movements with detector arrays, and other novel detector systems, is described in Section 2.8.1. These data and appropriate models can provide information on key parameters of salmon life history including inter-stage survival rates, migration rates and behaviour.

A number of recent publications have used these technologies to address questions of marine mortality from estuarine and nearshore waters to large bays. As reported in ICES (2012a), Kocik et al. (2009) describe smolt migrations on the Narraguagus River (Maine, USA) based on six years of monitoring (1997-1999 and 2002-2004). Migrating smolts were captured in the lower river, surgically tagged and released. Receivers were deployed throughout the lower river, estuary and nearshore environment to estimate smolt survival to the Gulf of Maine, map migration paths and document emigration timing for this population. Estimates of overall survival indicated that for every 100 smolts exiting the river, 62-74 reached the Inner Bay, 41-54 reached the Middle Bay, and 36-47 reached the Outer Bay. While mortality decreased in the marine environment, analysis indicated that less than half the smolts survived the approximately ten day period of migration from the river to the Gulf of Maine.

Dempson et al. (2011) reported on a three year initiative to track Atlantic salmon and determine migration route, residency time and survival in a 50 km long estuarine fjord located on the south coast of Newfoundland, Canada. Migrating smolts from two rivers in the study used different routes to reach the outer areas of the fjord. Many smolts were resident for periods of $4-8$ weeks moving back and forth in the outer part of the fjord where maximum water depths range from 300 to 700 m . Survival in the estuary zone was
greater for smolts with prolonged residency in estuarine habitat. Overall smolt survival to the fjord exit was moderately high (54-85\%), indicating that the initial phase of migration did not coincide with a period of unusually high mortality.

Halfyard et al. (2012) used acoustic tracking technologies to estimate mortality rates, assess the spatio-temporal dynamics of natural mortality and examine migratory behaviour during the fresh to saltwater transition of Atlantic salmon smolts from four river systems in the Southern Uplands area of Nova Scotia (Canada). They reported that the cumulative survival through the river, inner estuary, outer estuary and bay habitats averaged $59.6 \%$ (range $=39.4-73.5 \%$ ).

Lacroix (2013) reported on research to describe the migration of wild and hatchery Atlantic salmon postsmolts across the Bay of Fundy (BoF), Canada. This followed on previous publications (Lacroix et al. 2005; Lacroix 2008) to estimate survivals of smolt and postsmolts from the same area.

## Modelling detection and survival probabilities

Kocik et al. (2009) and Dempson et al. (2011) estimated detection and survival probabilities of tagged smolts using a variant of a Cormac-Jolly-Seber capture and recapture model in a program called MARK (White and Burnham, 1999). Lacroix (2008) and Halfyard et al. (2012) estimated detection probabilities independently of the tagged smolts and subsequently estimated survival probabilities outside a formal model structure.

An example of modelling individual fish detection data using smolts tagged from three rivers and six years and detected at arrays extending more than 800 km from the point of release was provided to the Working Group. A Bayesian state-space model variant of the Cormac-Jolly-Seber model described by Gimenez et al. (2007) and Royle (2008) was used. This model provides a means of disentangling the imperfect detection (p) of tagged smolts on the sonic arrays from apparent survival ( $\varnothing$ ) during their out migration. The state process (survival) is represented by a binary variable $z(i, j)$, which takes the value 1 if fish $i$ is alive at the end of the $j$ migration leg and 0 otherwise, with a probability of survival ( $\varnothing \mathrm{j})$ that is similar for individual fish within a stage of migration $j$.

$$
z[i, j] \mid z[i, j-1], \varnothing_{j} \sim \operatorname{Bernoulli}\left(z[i, j-1] \varnothing_{\mathrm{j}}\right)
$$

The observation process (detection) is also modelled by a binary variable, where $x(i, j)$ (observation) represents fish $i$ being detected at array $j$ conditional on $z(i, j)$ (i.e. whether the fish is alive to be detected at array $j$ ) and the probability of detection ( $p_{j}$ ) at the array for the migration leg $j$.

$$
x[i, j] \mid z[i, j], p_{j} \sim \operatorname{Bernoulli}\left(z[i, j] p_{j}\right)
$$

In the first analysis, an annual model was considered in which the probabilities of detection and survival were assumed independent among rivers, years, and detection arrays (Figure 2.3.4.3.1). In the second analysis, a hierarchical structure was placed on the probabilities of detection at the arrays. In the hierarchical model, the probabilities of detection were considered exchangeable among years within a river, and among years and rivers for those which share a common bay or exit farther out at sea.

Data collected by the Atlantic Salmon Federation of smolts tagged from three rivers and six years and detected at three arrays extending more than 800 km from the point of re-
lease were used to illustrate how these data could be analysed using the Bayesian statespace model formulation. A total of 1279 smolts had been tagged and released with acoustic transmitters from three rivers over a period of six years. Acoustic arrays were monitored at the head of tide of each river, at the exit to the Gulf of St Lawrence (two outer arrays) and at the Strait of Belle Isle (one array) leading to the Labrador Sea. The detections of smolts at each array by river and year are shown in Figure 2.3.4.3.2.

The estimates of the probabilities of detection at each array for the annual model and for the hierarchical implementation of the Bayesian state-space model are shown in Figure 2.3.4.3.3. The most relevant features from these results are:

- There is large annual variation in the probabilities of detection among years within a river at the head of tide arrays and among years at arrays exiting to the Gulf of St Lawrence.
- The uncertainties in the estimation of probabilities of detection increase as the monitoring proceeds from the head of tide arrays, to exit to the bays to the Strait of Belle Isle. This is due mostly to the smaller number of detections of smolts at the progressively further downstream arrays associated in part with fewer number of tagged smolts available for detection (fish die over time).
- The probability of detections at the last array at the Strait of Belle Isle are confounded with the probabilities of survival through the Gulf of St Lawrence and cannot be appropriately estimated. The range of detection probabilities vary from a high of 1.0 (perfect detection) assuming that the total fish detected corresponds to the total fish surviving the migration from the exit of the bays to lows corresponding to the ratio of detections relative to the estimated number of smolts alive at the previous detection array ( $100 \%$ survival through the Gulf of St Lawrence).
- Assuming some degree of exchangeability between the probabilities of detection results in shrinkage (reduced uncertainty) and slight changes in expected values of the annual and river-specific detection probabilities at the head of tide and bay exit arrays. The posterior distribution of the Strait of Belle Isle array under the hierarchical distribution is the mean of the individual year and river distributions and the confusion between the probabilities of detection and survival persists.

The posterior distributions of the probabilities of detection at the Strait of Belle Isle array are entirely determined by the prior assumptions for this parameter (Figure 2.3.4.3.4). However, the probabilities of detection and the estimates of survival at all the prior/upstream arrays are insensitive and unaffected by the estimates of the probability of detection at the last array.

Experience from the last ten years of research with the use of acoustic technologies to track salmon smolts provides useful guidance in the design of such experiments and the treatment of data.

- The last array, in time and/or space is the weakest point in the experimental design. It is not possible to disaggregate the probabilities of detection from the probabilities of survival unless an informative prior is used for this parameter
in the model or sampling efforts to detect tagged fish are expended downstream/later in time of the last array of interest. An informative prior could be developed by independently determining the probabilities of detection using sentinel tags (tags placed or transported across various parts of the array) as was done by Halfyard et al. (2012). Similar work has been initiated for the Strait of Belle Isle array.
- There will inherently be more uncertainty in estimating survival rates through the extended period of migration of salmon in the ocean as the sample size will decrease over time as fish die and fewer fish remain to be detected.
- Bayesian hierarchical models provide a flexible framework for analysing multiyear, multiarray, and multiriver designs. Bayesian models are flexible and additional variables can be introduced to further explore factors modifying detection and survival, for example by tag release group or date of release, by size of smolt, by incorporating indices of potential predators, etc.

In this example, hierarchical estimates of detection probabilities for the exit to the Miramichi Bay arrays can be used as informative priors to estimate survival rates to exit to the Gulf of St Lawrence in years when the Strait of Belle Isle array was not installed and operating (2004 to 2006).

### 2.3.5 The impact of artificial night light on Atlantic salmon fry dispersal and the onset of smolt migration

The use of artificial night light is continuing to increase both in previously unlit regions of the developing world, but also in already heavily developed countries. Different types of lights have varying spectral compositions. The most numerous current type of street lights emit light that is narrowly concentrated in the longer wavelengths of the visible spectrum, appearing yellow or orange to the eye. Modern, replacement, lights emit considerably more light across the visible spectrum especially at shorter wavelengths, providing high efficiency and superior colour rendering for human vision. However, these more natural whiter lights could lead to significant changes in the impact of artificial light on natural systems, particularly aquatic ecosystems where penetration through water will be increased.

In a recent investigation, the timing of Atlantic salmon fry dispersal from artificial redds (Riley et al. 2013; Riley et al. in prep); and the migratory timing and behaviour of wild smolts leaving their natal stream (Riley et al. 2012); were compared under both control and ecologically relevant broad spectrum street-lit intensities. The diel timing of both behaviours is considered to be a predator avoidance tactic for these critical life-history stages. Thus, any alteration or disruption to these processes may have a significant impact on recruitment.

Two experiments were conducted on fry dispersal, in 2011 and 2012. In the first experiment, fry dispersal occurred 2.8 days later ( $p<0.001$ ), and on average the fry were smaller at dispersal ( $p<0.001$ ), in the incubators exposed to a 12 lx light intensity compared to the 0.1 lx controls (Figure 2.3.5.1). In the second experiment, fry dispersal occurred 1.3 days later ( $p<0.01$ ) in incubators exposed to a 1.0 lx intensity compared to the 0.1 lx controls. In both experiments, significant disruption to the diel pattern of fry dispersal was also observed. For example, in 2011 fry dispersal under control conditions was significantly
directed around a mean time of $4: 17 \mathrm{~h}$ after dusk ( $p<0.001, \mathrm{r}=0.76, \mathrm{n}=1990$ ) with very few fry ( $<2 \%$ ) dispersing during daylight hours (Figure 2.3.5.2b). Under street lighting, the dispersal of fry was significantly delayed (mean time 6:38 h after dusk; $p<0.001, \mathrm{r}=$ $0.39, \mathrm{n}=2413$ ) with a significant proportion (32\%) dispersing during daylight hours (Figure 2.3.5.2a).

The migratory timing of smolts leaving their natal stream was determined using a PIT antennae system on the Brandy Stream, a tributary of the River Itchen in UK (England \& Wales). Experiments compared the downstream migration of smolts under natural control conditions (2000-2006) with two years (2008 and 2009) when the downstream exit to the study site was subject to street-lit conditions every alternate night (maximum light intensity = 14 lx ). Migration of smolts under control conditions was significantly ( $p<0.01$, $\mathrm{n}=170$ ) correlated with sunset. By contrast, street lighting resulted in the timing of migration being random ( $p=0.11, \mathrm{n}=7 ; p=0.76, \mathrm{n}=34$, respectively) with respect to time of day. Furthermore, migration of smolts was significantly ( $p=0.01, \mathrm{n}=19$ ) correlated with the time of sunset for fish migrating when the lamp was off, but random $(p=0.36, \mathrm{n}=22)$ when the lamp had been on (2008 and 2009 data, combined) (Figure 2.3.5.3).

Systematic investigation is needed to determine the extent of this problem and the light intensities at which street lamps do not affect behaviour. Such information could then be used as a management tool to identify sites where potential problems currently exist and provide evidence-based information to guide the replacement of street lamps to lessen their impact.

### 2.3.6 Stock identification of salmon caught in the Faroes salmon fishery

Salmon originating in rivers from both Northern and Southern European stock complexes have been exploited in the longline fishery that operated within the Faroes EEZ in the 1980s and 1990s, and there is a potential for this fishery to reopen if stocks recover. NASCO has asked ICES to develop a risk-based framework for the provision of catch advice for this fishery (Section 3.4), but this has been complicated by lack of data on the stocks that the fishery exploits. Advances in microsatellite DNA profiling methodologies and statistical genetics approaches, including work undertaken under the SALSEAMerge programme, provide the opportunity to obtain estimates of the stock composition in the fishery area for one or more baseline years.

Preliminary results were reported from a genetic study of salmon scales collected in the Faroes salmon fishery in the 1980 and 1990s. This study involves scientists from UK (Cefas and Marine Scotland Science), Norway (NINA and IMR) and Faroes (MRI) and is funded by the NASCO IASRB, and by UK, Norwegian and Irish government departments.

Approximately 750 scale samples collected from commercial and research catches in the fishery were selected from each of two periods comprising the 1983/1984 and 1984/1985 seasons and the 1991/1992 and 1992/1993 seasons respectively. Initial results have shown significant degradation of the DNA in some of the monthly samples, but much better results in others. Further investigations are being undertaken into the cause of the degradation, and initial trials with modified protocols suggest that it may be possible to improve the extraction of useable DNA. Although no assignment analysis has been
undertaken yet, a number of samples have been identified with alleles that are only expected to occur in North American salmon.

### 2.3.7 Update on EU project ECOKNOWS

ECOKNOWS is an EU 7th framework project running from 2009 to 2014, comprising of 13 research organizations with the University of Helsinki (Finland) leading. The project is to develop methodologies using Bayesian approaches. Developments are demonstrated in case studies, one of which is the salmon case study. The participants of the salmon case study are Agrocampus Ouest (France), the Finnish Game and Fisheries Research Institute, Institute National de la Recherche Agronomique (France), the Marine Institute (Ireland) and the Department of Fisheries and Oceans (Canada). In this case study the salmon stock assessment models used in the Baltic (in WGBAST) and North Atlantic (in WGNAS) areas are being compared with the aim of harmonizing the two approaches into comparable structures, mathematically representing salmon life cycles with freshwater and sea age cohorts. Both approaches are being developed with emphasis on improving the use of ecological knowledge and available data in assessments and improving the predictive ability of models.
Models are being developed for North Atlantic salmon stocks that have the potential to provide improvements to the Pre Fishery Abundance stock assessment models. An integrated life cycle model has been developed in a Hierarchical Bayesian framework that brings a substantial contribution to Atlantic salmon stock assessment on a broad ocean scale. The approach also facilitates the harmonization of the stock assessment models used in the WGBAST and in WGNAS. One of the main deliverables will be progress towards embedding Atlantic salmon stock assessment at broad ocean scale within an integrated Bayesian life cycle modelling framework consisting of two main components, as outlined below.

### 2.3.7.1 An integrated life cycle model as an improvement to PFA modelling

A life cycle model has been developed in a Hierarchical Bayesian modelling framework. The existing biological and ecological knowledge of Atlantic salmon demography and population dynamics is first integrated into an age and stage-based life cycle population dynamic model, which explicitly separates the freshwater (egg-to-smolt) and marine phases (smolt-to-return), and incorporates the variability of life histories (river and sea ages). The marine phase accounts for natural and fishing mortality, and captures the sequential fisheries along the migration routes, including off shore, coastal and estuarine and freshwater fisheries. This body of prior knowledge forms the prior about the population dynamic, which is then updated through the model with assimilation of the available data.

The framework offers potential improvements to the PFA stock assessment approach. The current PFA models mainly rely on a stock-recruitment concept that considers a statistical relationship between a spawning potential (lagged eggs) and a recruitment variable (PFA), both derived from the same data sources (estimates of returns based on home water catches) by a mixture of forward (lagged eggs) and backward (run reconstruction) approaches. The freshwater phase is not explicitly represented in the model. More generally, many demographic hypotheses are obscured within the data assimila-
tion procedure, making it difficult to assess how changes in models or data may impact the results.

The new modelling approach makes it easier to assess the consequences of any changes in the data and model structure. Different demographic hypotheses can be tested without changing the data assimilation scheme, and this would also offer multiple possibilities to extend the model by adding more sources of data (e.g. data on egg-to-smolt survival, post-smolt mark-recapture data, environmental variables). As a critical improvement to the PFA models, the life cycle model explicitly separates out the freshwater and the marine phases. This allows the effect of both freshwater and marine phases in the recruitment process to be separated, instead of considering a single productivity parameter that aggregates demographic processes of the different impacts encountered during the freshwater phase (from egg to smolt) and the first months of the post-smolt marine phase.

To illustrate the potential of the approach, the model has been applied to the stock complex of Eastern Scotland, the largest regional component of the southern NEAC stock complex. The model was fitted to the same data as used in the current PFA models. In addition to the hypotheses currently made in the current PFA model, the flexibility of the approach has been illustrated by testing different demographic hypotheses.

- Density-dependence in the freshwater phase has been considered by introducing a Beverton-Holt egg-to-smolt survival (Figure 2.3.7.1.1). This introduced non-linearity in the dynamics and modified the inferences made on the smolt-to-PFA survival (marine productivity). Indeed, the fluctuations in number of eggs spawned over the time-series induced fluctuations in the egg-to-smolt survival rate, which is balanced by changes in the smolt-to-PFA survival relationship. Hence, considering density-dependence in the freshwater phase leads to a different time-series of egg-to-smolt survival estimates. This in turn may change interpretation of changes in marine productivity and may also affect forecasts.
- The model was also used to contrast two hypotheses for the decline in return rates of 2SW fish (Figure 2.3.7.1.2): a constant natural mortality rate after the PFA stage and an increase in the proportion maturing (current hypothesis in PFA models); or an increase in the natural mortality rate of 2 SW fish relative to 1SW, and a constant proportion maturing. Changing from one hypothesis to the other has no consequence on estimates of smolts return rates, but it supposes different demographic processes. It may also have management implications as a greater mortality on 2SW fish would result in a different risk to stock abundance in homewaters from high seas fisheries.

Proposed further work includes: (i) building a hierarchical model to jointly analyse the dynamics for all regions comprising the southern NEAC stock complex; (ii) enhancing the validation of the available data. Such a model has the potential to improve knowledge of the biology and ecology of Atlantic salmon. In particular, future development will consist of including region specific egg-to-smolt productivity parameters derived from meta-analyses based on index rivers (Section 2.3.7.2).

Such a model has the potential to provide tools for assessing the effect of management measures on mixed-stock high seas fisheries;

### 2.3.7.2 A meta-analysis of egg-to-smolt survival

A meta-analysis of egg-to-smolt relationships for Atlantic salmon has been carried out. Time-series of egg-to-smolt data on 21 index rivers across the Atlantic salmon range (12 rivers from North America; nine rivers in Europe), together with several covariates associated with the index rivers, were compiled for this study:

- Total egg deposition for each cohort, derived from estimates of the number of returning spawners combined with estimates of proportion of sea age classes, proportion of females and fecundity of each sea age class;
- The total smolt production by cohort including age-structure of the smolts;
- Associated covariates for each river: latitude (continuous), longitude (categorical with two groups, east and west side of the Atlantic Ocean), wetted and lacustrine area accessible to salmon.

The meta-analysis was carried out through a Hierarchical Bayesian model. The classical Beverton-Holt model was revisited through the explicit parameterization in terms of density-independent and density-dependent mortality rates. The duration of the freshwater phase (mean age of smolt, specific to each river) is explicitly used as a covariate. A partially exchangeable hierarchical model was built to incorporate covariates (such as the longitude and the latitude) to capture part of the between rivers variability.

Results highlight large between rivers variability in both the density-independent and dependent mortality rates (Figure 2.3.7.2.1). Latitude and the longitude explain a great part of this variability in the density-dependent mortality rate (Figure 2.3.7.2.1). No useful covariates were found however, to explain the variability in the density-dependent mortality rates. This approach offers an efficient framework to predict the parameters of density-dependent survival (and the associated uncertainty) for any new river for which the associated covariates of latitude, longitude, wetted area and mean smolt age are known.

Outcomes of this study offer useful prior information about freshwater productivity, which may be worked into the integrated life cycle model described above which is being developed in parallel with this process.

### 2.3.8 Diseases and parasites

### 2.3.8.1 Red vent syndrome

Over recent years, there have been reports from a number of countries in the NEAC and NAC areas of salmon returning to rivers with swollen and/or bleeding vents. The condition, known as red vent syndrome (RVS or Anasakiasis), has been noted since 2005, and has been linked to the presence of a nematode worm, Anisakis simplex (Beck et al. 2008). This is a common parasite of marine fish and is also found in migratory species. However, while the larval nematode stages in fish are usually found spirally coiled on the mesenteries, internal organs and less frequently in the somatic muscle of host fish, their presence in the muscle and connective tissue surrounding the vents of Atlantic salmon is unusual. The reason for their occurrence in the vents of migrating wild salmon, and whether this might be linked to possible environmental factors, or changes in the num-
bers of prey species (intermediate hosts of the parasite) or marine mammals (final hosts) remains unclear.

A number of regions within the NEAC stock complex observed a notable increase in the incidence of salmon with RVS during 2007 (ICES, 2008a), but levels have been lower in some NEAC countries since 2008 (ICES, 2009a; ICES, 2010b; ICES, 2011b). Trapping records for rivers in UK (England \& Wales) and France suggest a further reduction in 2012. The incidence of Anisakis simplex was considered to be much lower in 2012 compared to the previous four to five years in Ireland.

There is no clear indication that RVS affects either the survival of the fish or their spawning success. Affected fish have been taken for use as broodstock in a number of countries, successfully stripped of their eggs, and these have developed normally in hatcheries. Recent results have also demonstrated that affected vents showed signs of progressive healing in freshwater, suggesting that the time when a fish is examined for RVS, relative to its period of in-river residence, is likely to influence perceptions about the prevalence of the condition. This is consistent with the lower incidence of RVS in fish sampled in tributaries or collected as broodstock compared with fish sampled in fish traps close to the head of tide.

### 2.3.8.2 Monitoring of sea lice burdens on wild returning adult Atlantic salmon

There remains a paucity of studies of sea lice prevalence and intensity on Atlantic salmon in areas prior to the development of aquaculture and in areas currently without aquaculture. Powell et al. (1999) reported on prevalence and abundance of sea lice on migrating Atlantic salmon monitored at a fishway near the head of tide in the Penobscot River (USA), as the aquaculture industry was developing in the area. Bjørn et al. (2001) report on the prevalence and abundance of sea lice from Atlantic salmon sampled from catches in various coastal and inshore fisheries in Norway from June and July 2001. Prevalence of lice on salmon ranged from $80 \%$ to $100 \%$ and the maximum numbers of lice on fish among locations ranged from 28 to over 1000 lice. Murray and Simpson (2006) reported on monitoring of sea lice from salmon in the estuary fishery of the North Esk river (UK Scotland) during 2001 to 2003. This river is distant from marine salmon farms and provides a source of data on background patterns in lice infestation on returning salmon. In this study, mobile lice had mean abundance of six to seven per fish and a prevalence of $80-90 \%$. Jackson et al. (2013) reported on population structure, prevalence and intensity of lice from salmon sampled from the driftnet fishery and draft net fishery catches in Ireland. The authors noted that almost all fish examined in these fisheries had sea lice with abundance variable both within and between years with the maximum mean abundance of 25.8 lice per fish recorded in 2004. There was clear evidence of recent infestation with lice in the draft net samples (estuarine areas) (Jackson et al., 2013).

## Monitoring of sea lice burdens on wild returning adult Atlantic salmon from the Miramichi River, New Brunswick

Results from a monitoring program from 2005 to 2011 that developed indices of sea lice abundance on returning Atlantic salmon to the Miramichi River, southern Gulf of St Lawrence (Canada) were presented. There is no marine finfish aquaculture in the southern Gulf of St Lawrence. Salmon were captured at research and monitoring estuary trapnets operated during the entire migration period for salmon (late May to late October).

Captured fish were sampled for biological characteristics including origin (wild vs. hatchery), length, scales for age determination, sex, sea lice scarring, and sea lice load. Monitoring for sea lice began in 2005. Sea lice on salmon were enumerated in five categories of abundance ( $0,1-5,5-15,15-50$, and $>50$ lice). Sea lice prevalence (percentage of fish with sea lice present) was lowest in June, increased over summer to generally highest levels in August although in a few years, the percentage of salmon with sea lice was higher in September (Figure 2.3.8.2.1). Sea lice loads, expressed as the percentage of fish in the $>15$ sea lice category, were highest in September with as many as $5 \%$ of the sampled fish having greater than 50 lice per individual fish in some years.

Sea lice infection rates on returning wild adult salmon increase through summer and autumn, which could be explained by Atlantic salmon collectively returning to Miramichi Bay in late spring/early summer and staging in Miramichi Bay with salmon ascending the river in late summer and autumn having been exposed for a longer period of time to sea lice in a constrained area.

Sea lice have not been identified to species so the observations could be of either Lepeophtheirus salmonis or Caligus sp. Despite the absence of salmonids in the brackish and saltwater portions of the Miramichi River and bay which freeze over in winter, sea life cycles are maintained in this area. Sea lice infections on returning wild adult Atlantic salmon can be quite high in some years and increase through summer and autumn. This indicates a "natural" state of the association between the ectoparasite and Atlantic salmon in an area without marine salmonid aquaculture.

## Monitoring of sea lice burdens on Atlantic salmon from the fishery at West Greenland

As part of the Enhanced Sampling Program (SALSEA Greenland; ICES 2012) information on the prevalence of sea lice on Atlantic salmon has been collected. Enhanced sampled fish provide a more unbiased estimate of sea lice prevalence at West Greenland than the Baseline sampled fish. Enhanced sampled fish were purchased directly from the fishers whereas fish sampled at the local market were sometimes cleaned prior to sampling at the local market, thereby removing most lice present. The fish samples from the Enhanced Sampling Programme at West Greenland may still be biased as sea lice are sometimes removed from individual fish due to abrasion against the gillnet during capture. As a result sea lice estimates from Greenland harvested fish should be considered minimum estimates. Samplers were instructed to document the presence and number of sea lice on each fish sampled. Subsamples of individual sea lice were also preserved in support of two ongoing studies: a Slice ${ }^{\circledR}$ resistance study and a population genetics study.

Information of sea lice prevalence is available from 1166 fish sampled between 2009 and 2011. On average, $30 \%$ of the sampled fish had no sea lice present (Table 2.3.8.2.1). Approximately $50 \%$ of the individuals had $1-5$ lice, $12 \%$ had $6-10$ lice fish and the remainder had eleven or more. Sea lice burden per fish ranged from 2.3 (2011) to 3.0 (2009 and 2010) with an overall burden of 2.7 lice per fish.

## Summary and considerations for improving the monitoring

The principal concern for sea lice originating from aquaculture relates to the impact of lice on outmigrating post-smolts which are most susceptible to these infections. It is challenging, but not impossible, to sample smolts and early post-smolt stages as they migrate
to the open ocean. Monitoring of sea lice burdens on adult salmon returning to rivers could be an alternate indicator of variations in abundance of sea lice among areas and among years. However, returning adults may be more indicative of the sea lice infestations in high seas than the sea lice infection pressure experienced by the outmigrating smolts.

Although sea lice infection rates can vary among locations due to differences in biological and oceanographic conditions, monitoring of sea lice infection rates on salmon populations in areas with and without salmon farms would provide information on the relative roles of salmon farms as a source of sea lice for wild salmonids. The "natural" state of the association between sea lice and Atlantic salmon in areas without marine salmonid aquaculture could be useful indicators of how these associations vary with factors unrelated to concerns about aquaculture. Sea lice development is temperature dependent and variations in lice loads on salmon may reflect variations in generation time for lice among different areas. The identification of the species and the life stage is important and in most studies, motile life stages are counted and mature females with egg cases are tabulated separately. Monitoring protocols have been developed by state agencies and industry and training courses for sea lice monitors are mandatory in some areas.

### 2.3.8.3 New parasite

In 2011, a parasite (Paragnathia formica, an estuarine crustacean isopod) was detected on $5 \%$ of salmon caught in the Scorff trap facility, France, located near the upper limit of the estuary. It was not clear whether this was a new infestation or one that had simply gone undetected until that point. Symptoms included inflammation in the genital zone and on the fins and could be mistaken for sea lice bites or red vent syndrome. It was also noted that sea lice and Paragnathia formica could be coexisting, and not known whether the parasite could survive in freshwater. Paragnathia formica was not detected in France in 2012.

### 2.3.9 Changing biological characteristics of salmon

The Working Group noted that various biological characteristics of salmon show continuing interesting trends, some of which have already been reported in the ICES SGBICEPS report (ICES, 2010a), such as decreasing mean fork lengths in returning adult 1SW fish in the River Bush in UK (N. Ireland) since 1973. This same trend has been observed for 1SW returning adults on the River Bann in UK (N. Ireland). For the same time period the mean fork length in 2SW fish in both rivers showed only a very small, nonsignificant, decrease.

Also notable, and first reported in ICES, 2011a, was the increase in both numbers of 2SW returns to the River Bush in UK (N. Ireland) as well as the increase in the relative proportion of 2 SW vs. 1SW, since 2003. In 2012 this trend has become even more pronounced with the percentage of River Bush 1SW returning adults decreasing to the lowest point in the time-series at $66 \%$ (previous ten year average $91 \%$ ). Survival to freshwater of River Bush 2SW fish has also seen a positive trend since 2001. This was also noted in the Norwegian PFA estimates for 1SW, 2SW, and 3SW returning adults. From the 2004 smolt cohort onwards, the estimates for 1 SW fish have decreased to approximately $15 \%$ and have remained low. PFA estimates for 2 SW and 3 SW returning adults for the same period have shown an opposing trend with a 10-20\% increase from 2004 (3SW) or 2005 (2SW). Angling catches in UK (England \& Wales) have noted a marked increase in the
proportion of $2 S W$ salmon relative to $1 S W$ salmon in the last two years. The above observations could indicate a shift in life-history strategy from 1SW to MSW in some Northern NEAC and Southern NEAC stocks, possibly due to poor growth in the first season at sea. A similar observation has been noted in the probability of maturing parameter in the stock and recruitment model the Northern NEAC model (Section 3.5).

Another possible change in biological characteristics has been observed in mean smolt age in UK (England \& Wales). In rivers such as the Dee there has been a downward trend in smolt age since the late 1960s (ICES, 2010a) and similar decreases, over shorter timescales, were observed in other rivers in the same geographic area. However, since 20032004 this trend appears to have been reversed and mean smolt age on the Dee, and other monitored rivers, appears to be increasing again.

### 2.3.10 New initiatives in relation to management of mixed-stock coastal fisheries in northern Norway

SALSEA-Merge, and other projects, have contributed towards the establishment of a comprehensive genetic baseline for salmon populations in northern Europe. This baseline continues to be developed as a practical and useful tool for management of mixed-stock coastal fisheries in Norway and Russia. In 2010 the Working Group reported (ICES, 2010b) on a pilot project that expanded the baseline for a number of Russian rivers, and ongoing genetic analysis and assignment of samples from salmon caught in coastal fisheries in Norway. Power analysis of the genetic baseline developed indicated that with the baseline coverage, and number of genetic markers used, approximately $50 \%$ of the samples from coastal fisheries can be reliably assigned to river (probability $>90 \%$ ). A total of 1900 samples from adult salmon caught in coastal fisheries in 2008 in Finnmark county, northern Norway, were genetically analysed and assigned to defined geographical regions or rivers in the baseline (Svenning et al. 2011). The results demonstrated that the applied method can give reliable estimates of the proportion of Russian salmon in the catches as well as estimates of how salmon from different regions are exploited in the coastal fisheries (see Section 3.1.8).

However, it was also recognized that the spatial coverage of the baseline should be expanded, the number of genetic markers should be increased, and additional sampling should be conducted in a number of salmon rivers in Norway and Russia to improve the precision of the assignment of individuals. A further initiative to achieve this goal has been taken by Norway, the Russian Federation and Finland. In 2011 a new EU project "Trilateral cooperation on our common resource; the Atlantic salmon in the Barents region" (Kolarctic Salmon) was started. The project is supported by both EU-funding (Kolarctic ENPI CBC) and national funding from Norway, the Russian Federation and Finland.

In 2011 and 2012, the genetic baseline was expanded both in terms of spatial coverage and completeness, and it now contains genetic data from over 180 salmon populations in northern Norway, Finland and Russia. The number of genetic markers has been upgraded to 31 microsatellite loci. Over 17000 samples were collected from coastal fisheries in northern Norway and Russia in 2011 and 2012, and analysis of these samples is now underway. Preliminary assignment of a subset of these samples has already provided valuable information on the composition of catches in time and space, and interesting
patterns of coastal migration of different populations and sea age groups are beginning to emerge.

Moreover, the potential use of other genetic markers - single nucleotide polymorphisms (SNPs) - for genetic stock identification in Atlantic salmon was evaluated (Ozerov et al. 2013). This work demonstrates the possibility of cost-effective identification of dozens of informative SNPs (among thousands) for discrimination of populations at various geographical scales, as well as identification of loci controlling ecologically and economically important traits. Through the activities in this project, a foundation will be established on which a river-specific management regime for coastal and riverine fisheries for these northern populations can be implemented.

### 2.4 NASCO has asked ICES to provide a review of examples of successes and failures in wild salmon restoration and rehabilitation and develop a classification of activities which could be recommended under various conditions or threats to the persistence of populations

The Working Group on the Effectiveness of Recovery Actions for Atlantic salmon (WGERAAS) met for the first time in Belfast, UK (Northern Ireland), from the 18th to the 22nd of February 2013. The meeting was attended by 22 delegates from eleven countries. The ToRs were as follows:

1 ) Develop a classification system for recovery/re-building programmes for Atlantic salmon, including threats to populations, population status, life-history attributes, actions taken to re-build populations, program goals, and metrics for evaluating the success of re-building programmes.
2 ) Populate the system by collecting data on recovery/re-building programs for Atlantic salmon populations from around the North Atlantic.
3 ) Summarize the resulting dataset to determine the conditions under which various recovery/re-building actions are successful and when they are not.
4 ) Provide recommendations on appropriate recovery/rebuilding actions for Atlantic salmon given threats to populations, status and life history.

WGERAAS concluded that the most appropriate way to address the first and second TOR is to develop a database which lists threats to populations, population status, actions taken to rebuild populations, at river level. WGERAAS recommended this database be established through an update of the NASCO River Database by adding additional columns to this database. These columns will consist of: (1) 'population status' (this field already exists in the current version, but could usefully be updated), (2) ten columns of population 'stressors' or threats, and (3) ten columns of recovery actions. These columns will feature a drop-down menu with a limited choice of answers. For the 'stressors' columns for example, these answers will range from 'Very Strong' to 'Unknown'; the default (no information) will be to leave the field blank. WGERAAS recognized that these data would be best provided by regional or national experts. A guide on how to fill in this database (including examples) is being developed and will be provided to those people asked to populate the database with data. The data in the database will be used to provide a broad perspective of the scale of different stressors and recovery actions being
applied around the North Atlantic. To address the third ToR these data will be analysed, and together with case-studies on the effectiveness of different recovery and restoration actions, discussed in the final report. From this discussion the conditions will be determined when recovery/rebuilding actions are successful, and when not. With this information the fourth ToR can be addressed i.e. producing recommendations on the appropriate recovery/rebuilding actions for Atlantic salmon populations. Therefore it was proposed that WGERAAS approach NASCO to allow the Group access to and utilize the existing NASCO rivers database.

In addition to the proposal outlined above, WGERAAS noted that a meta-analysis determining which factors influence the success of recovery and rebuilding actions for Atlantic salmon, could potentially prove informative in further evaluating the effectiveness of recovery actions and providing a broader perspective of the actions taken across the full range of the species. A more detailed questionnaire than the updated NASCO Database would be necessary for a high quality meta-analysis. Currently, WGERAAS sees no opportunity to conduct this meta-analysis within the group. A possible complementary initiative is currently being explored by several members of the WGERAAS group. A proposal has been submitted for an investigation supported by the EU entitled SALAWARE (Safeguarding our Atlantic Salmon Cultural Heritage-Europe's Oldest Natural Legacy). Such external funding could hopefully support a PhD or Post-Doc to conduct a more detailed analysis. WGERAAS is scheduled to meet again in January 2014 at ICES in Copenhagen. The Workshop on Wild Atlantic Salmon Recovery Programs (hosted by the Atlantic Salmon Federation) and scheduled for September 18th and 19th 2013 in St Andrews, New Brunswick, Canada, and a UK symposium on stocking (organized by the Atlantic Salmon Trust and The Loughs Agency and to be held in 27th and 28th November 2013 in Glasgow, Scotland) are also expected to inform WGERAAS deliberations. A final report is due before the 2014 meeting of WGNAS.

The Working Group on North Atlantic Salmon (WGNAS) welcomed the progress made by WGERAAS at its first meeting and noted that work to address the ToRs was at an early stage. WGNAS had some concerns that the time frame for WGERAAS to submit a final report might not be sufficient to gather all the data required to address the ToRs successfully. WGNAS suggests WGERAAS might want to consider an extra year to gather data and produce a final report. WGNAS also has some reservations regarding the database part of WGERAAS's chosen approach, specifically the scientific rigor of the method within and among contributors, potential issues with the quality and completeness of the answers, and how the data might be interpreted beyond the specific work of the WGERAAS. WGNAS suggests WGERAAS put more emphasis on developing the case-studies on recovery/restoration actions as a method for addressing the ToRs and in this regard offered to encourage the identification of case studies on recovery and restoration actions which may have occurred or are ongoing in their respective locations. The Working Group noted that NASCO has identified under its action plans that stock rebuilding programmes including habitat improvement, stock enhancement and fishery management actions should be considered for stocks which are below Conservation Limits (Section 1.4).

### 2.5 NASCO has asked ICES to advise on the potential threats to Atlantic salmon from exotic salmonids including rainbow trout and brown trout where appropriate

### 2.5.1 Introduction

The Working Group noted that salmonid species had been spread widely around the globe from their original native distributions. In particular, rainbow trout (Oncorhynchus mykiss) and different salmon species native to the Pacific have been introduced widely to other countries for use in aquaculture and in fisheries. Similarly, brown trout (Salmo trutta), native to Europe, have also been introduced widely to other countries, including North America. The extent to which these introduced species have established natural populations in new areas, or subsequently become invasive, has varied. However, concerns are often raised about the impact, or potential impact, which introduced species can have on native species and ecosystems. In considering this question on the potential threats from exotic salmonids, the Working Group noted that species translocated into waters where they previously didn't exist, but which fell within the biogeographical range of the species, might also pose a potential threat to established native populations, and thus be considered exotic.

The Working Group noted that the issue of threats to Atlantic salmon by introductions and transfers of salmonids had previously been reviewed by the NASCO North American Commission (NASCO, 1992) and that this included protocols for reducing the risk of ecological effects associated with such movements. The Working Group considered that the recommendations from this report remained valid.

### 2.5.2 Overview of current distribution of exotic salmonids

A broad overview of the current distribution and status of exotic salmonids in the main North Atlantic salmon producing countries is presented split into NEAC (Table 2.5.2.1a) and NAC (Table 2.5.2.1b) areas. Rainbow trout have been introduced throughout Europe and on the Atlantic coast of North America and are used extensively in both aquaculture and recreational fisheries. However, there are few records of the species establishing in NEAC areas. In contrast, Pacific pink salmon (Oncorhynchus gorbuscha) have been introduced and become established in Russia and parts of northern Norway. In the NAC area, brown trout introduced from Europe have become widely established and are spreading in many areas, and rainbow trout have also established in some areas. This section has focused largely on these three species; the other species referred to in the tables are not as widely distributed and only limited information was provided on these.

## Rainbow trout

Rainbow trout, Oncorhynchus mykiss, have been introduced to Eastern Canada from the Northern Pacific Ocean for recreational fishing. Stocking in the upper Saint Lawrence started in 1893 and is still authorized in some of these upstream regions. However, an increasing prevalence of rainbow trout in Atlantic salmon Rivers in Eastern Québec, Canada, has been observed since 1980s, although stocking is strictly forbidden in these water systems.

Rainbow trout are now present in about fifty river systems in Eastern Quebec and evidence of reproduction was found on twelve of them, suggesting the presence of selfsustained populations. According to genetic analyses, these individuals would come from naturalised upstream populations following stocking conducted in the upper SaintLawrence in the provinces of Quebec, Ontario and in the United States (Thibault et al. 2009). Otolith Sr:Ca analyses also revealed that, although all fish captured in the upstream stocking regions were freshwater residents, both anadromous and freshwater resident phenotypes were observed downstream in Eastern Quebec (Thibault et al. 2010a). In fact, the proportion of fish exhibiting an anadromous phenotype increased with the distance from the stocking zone, suggesting that the development of the anadromous life cycle enables this species to colonize new rivers following long distance migration in the Saint Lawrence Estuary (Thibault et al. 2010a).

According to a modelling analysis on the physical characteristics of the colonized river systems, the presence of rainbow trout is associated with the number of tributaries, warm spring and summer temperature, and negatively related to peak flood during egg deposition in May (Thibault et al. 2010b). The spreading of rainbow trout is a subject of concern for the Quebec government because it represents a strong competitor and predator, which can affect indigenous species in a water system (e.g. Coghlan Jr et al. 2007). In this context, the Ministère du Développement Durable, de l'Environnement, de la Faune et des Parcs (MDDEFP) of Quebec developed an Action Plan to improve knowledge, revise stocking practices, increase total catch, limit propagation, and inform and educate citizens concerning the status of rainbow trout in the province. This Action Plan is currently in press and should be published at the end of 2013.

Rainbow trout are not widely established in NEAC countries (Table 2.5.2.1.a). For example, there is currently thought to be only one small self-sustaining population in England despite widespread use of the species in aquaculture and in recreational (put-and-take) fisheries dating back for more than one hundred and fifty years. A recent review of the risk of invasion of rainbow trout in the UK (Fausch, 2007) has indicated that the species has generally failed to establish. The primary abiotic factors (e.g. temperature and flow) that commonly influence the success of many stream fishes are not thought to have been limiting. Rather, the factors considered most likely to be constraining establishment in the UK, alone or in combination, are: biotic resistance from native salmonids, parasites and diseases, and angling mortality. However, an important consideration in assessing potential threats to Atlantic salmon from exotic salmonids is that these can change over time. Thus, Fausch (2007) cautions that the current lack of establishment of rainbow trout does not equate to the absence of risk. The situation could change in future as a result of factors such as: climate change; other changes in environmental conditions; introductions of new strains of rainbow trout; declines in native salmonids; or other anthropogenic changes to river environments.

## Pink salmon

Pink salmon have the shortest life cycle among species of the genus Oncorhynchus, as they mature and reproduce after only two years. Therefore, there are two reproductively isolated populations spawning in alternate even and odd years (Heard, 1991).

In Russia, pink salmon were introduced to the White Sea basin in the 1950s with annual egg transfers from the Far East of Russia into hatcheries of Murmansk and Archangelsk
regions (Gordeeva and Salmenkova, 2011). Despite over 20 years of introductions no consistent natural reproduction occurred and they disappeared when the introduction stopped in 1979. This failure was attributed to use of populations from the southern part of the native range. As time of spawning migration and spawning time are strictly fixed in salmonids, the introduced "southern" pink salmon began to spawn too late and eggs were lost as water temperatures in Autumn were colder than in their native habitat especially in even-year generations (Dyagilev and Markevich, 1979). Therefore successful natural reproduction took place only during some years of the North Atlantic warming (Karpevich et al. 1991).

The introduction of odd-year pink salmon to the White Sea basin was undertaken in 1985, when a new broodstock population was selected from the northern part of the species range (Okhotsk Sea basin, Loenko et al. 2000). This single pink salmon egg transfer from an odd-year population resulted in the establishment of local self-reproducing populations in the White sea rivers of Murmansk and Archangelsk regions of Russia with the adult returns fluctuating between 60000 to 700000 fish during the period 1989 through 2009 (Zubchenko et al. 2004; Gordeeva et al. 2005). Pink salmon introduced to Russia since 1930s have resulted in catches in Norwegian waters (up to 20 t in some years). The species has also now established in eleven rivers in N. Norway (Finnmark); Hesthagen and Sandlund (2007). The commercial fishery for pink salmon takes place in the coastal areas of the White Sea and with the same gears and in the same season as Atlantic salmon fisheries. The total declared pink salmon catch in 2009 was 139 t , twice as much as a declared Atlantic salmon catch (ICES, 2010).

At the same time, transfers of even-year-broodlines from the same river of the Okhotsk Sea basin were unsuccessful despite the large number of eggs that were transferred and the favourable rearing conditions at hatcheries. The last egg transfer of 1998 resulted in comparatively large return in the first generation, but the abundance of pink salmon declined subsequent generations and after that they appeared only in small numbers in even years. No commercial fishery for pink salmon is conducted in the White Sea in even years.

## Brown trout

Brown trout are established in rivers in Newfoundland and Nova Scotia (2.5.1). Westley and Fleming (2011) looked at landscape factors that shaped the spread of brown trout in Newfoundland. Brown trout embryos were first shipped to Newfoundland from Scotland in 1883 and further importations continued until around 1906. The imported trout survived well and established populations in the watersheds surrounding St John's. It is believed that the trout escaped into watersheds with easy access to the sea around 1884 which provided a source of anadromous colonizers. Westley and Fleming (2011) concluded that brown trout had successfully invaded and established populations in watersheds in Newfoundland and that they were slowly expanding on the island. They also suggested that abiotic factors alone were not sufficient to prevent continued expansion and all watersheds in Newfoundland were potentially susceptible to successful invasion. Current distribution is estimated at 68 watersheds on the Avalon, Burin, and Bonavista peninsulas compared to 16 watersheds on the Avalon Peninsula in 1883. Westley et al. (2011) concluded that the mechanisms determining invasion success and failure remained largely unknown but that the outcome of interspecific competitive interactions
was highly context specific, varying among habitats, continents, and scales of investigation.

Few studies on the ecological impact of brown trout on native salmonids have been carried out in Canada. However impacts are believed to include competition and displacement of native fish (Gibson and Cunjak, 1986; Van Zyll de Jong et al., 2004) and hybridization with Atlantic salmon (Verspoor, 1988; McGowan and Davidson, 1992).

### 2.5.3 Potential threats posed by exotic salmonids

Non-native and translocated species can pose threats to native species and ecosystems in a number of ways. These include predation, competition, hybridization, and introduction of novel diseases and parasites. The main effect of these interactions is often a general displacement or replacement of native salmonids over time. The Working Group noted that information presented on the effects of exotic salmonids derived from a small number of country specific reports and are largely based on incidental findings and observations rather than directed studies. Westley et al. (2011) produced a review and annotated bibliography of the impacts of invasive brown trout (Salmo trutta) on native salmonids, with an emphasis on Newfoundland waters.

## Parasites and diseases

Exotic salmonids have the potential to transfer novel parasites or diseases to native Atlantic salmon populations. Rainbow trout have a high susceptibility to salmon lice (Lepeophtheirus salmonis) and where they co-occur also host parasites such as Gyrodactylus salaris.

The monogean G. salaris is a freshwater ecto-parasite of Atlantic salmon whose natural hosts are Baltic strains of Atlantic salmon. G. salaris was not found to cause host mortality on rainbow trout in Norway, but this species is a suitable host for the parasite, and capable of transmitting the parasite to new localities as a consequence of stocking programmes or migratory behaviour (Bakke et al. 1991). At present, G. salaris has been eliminated from all infected rainbow trout fish farms in Norway, and all farms producing rainbow trout in freshwater are inspected every two years for the occurrence of this parasite (Anon. 2011).

Rainbow trout may also disperse G. salaris between rivers through brackish water. Soleng and Bakke (1997) found G. salaris to survive and reproduce in $7.5 \%$, salinity for as long as 54 days at $12^{\circ} \mathrm{C}$. Few studies have examined the behaviour and spread of escaped farmed rainbow trout at sea, but they generally conclude that they disperse relatively slowly, and they prefer the warmer freshwater surface layer (Skilbrei, 2012). Jonsson et al. (1993b) concluded that rainbow trout were usually recaptured in the fjord area where they were released/escaped, and Skilbrei and Wennevik (2006) observed that the geographical distribution of gillnet recaptures of escaped rainbow trout agreed well with the localization of the fish farms and with escape events. Hence, this behaviour of rainbow trout in the fjord areas increases the risk for spread of G. salaris between rivers.

Rainbow trout have a high susceptibility to salmon lice (Fast et al., 2002; Gjerde and Saltkjelvik, 2009), and farming and escapees of rainbow trout may hence contribute to high infection rates of sea lice on wild salmonids. Holst (2004) observed a mean of 4.4 adult female sea lice on 115 rainbow trout captured with gillnets in late April/early May 1999 in
the Osterøy Fjord System, Norway. Considering that escaped farmed rainbow trout disperse relatively slowly, and hence occupy the same coastal area for long periods, these high infection rates suggest that they may contribute significantly to the production of sea lice larvae in the area from which they have escaped (Skilbrei and Wennevik, 2006). This risk is especially high for Atlantic salmon in areas where farms are located at smolt migration routes (Krkosek et al., 2009).

So far, no infections of other parasites or diseases have been reported in wild rainbow trout in Norway (Anon., 2011), although serious outbreaks of diseases like pancreas disease have been diagnosed in seawater fish farms (Taksdal et al., 2007; Kristoffersen et al., 2009).

Rainbow trout are suspected to be a host for G. salaris and in this regard future production of rainbow trout in aquaculture installations is of potential concern (Dagerman et al., 2012).

## Destruction of redds

Atlantic salmon spawn in autumn and the most common strains of rainbow trout spawn during spring. Thus, where rainbow trout do exhibit a degree of spawning behaviour, they may dig up and destroy salmon redds before the salmon fry emerge from the gravel. In Norway, rainbow trout have been filmed digging up redds of Atlantic salmon and/or brown trout (Anon. 2011). In Sweden, digging up of redds of brown trout by introduced rainbow trout has been considered a threat to brown trout populations (Landergren, 1999). In New Zealand, where rainbow trout populations have established, this behaviour has been an important mechanism in completely displacing brown trout populations (Hayes, 1987; Scott and Irvine, 2000).

Pink salmon migrate a shorter distance up rivers to spawn than most other salmonids (Heard 1991); in addition, spawning in pink salmon seems to be terminated before the spawning of Atlantic salmon starts. As such, there does not appear to be any evidence of interactions with Atlantic salmon at the spawning grounds, such as competition for spawning sites or destruction of redds.

Rainbow trout strains in North America may spawn in both spring and autumn so they may pose a higher risk to Atlantic salmon. While it is suggested that brown trout spawn prior to Atlantic salmon, redd superimposition is likely given that these species prefer similar spawning habitats (Heggberget et al., 1988; Louhi et al., 2008).

## Competition for territory, area and resources

Rainbow trout are considered to be generalist feeders, consuming a variety of food organisms. It has the potential to have a negative influence on habitat use and nutrient availability of native fish species (Elliott, 1973; Crowl, Townsend and Macintosh, 1992; Hasegawa and Maekawa, 2006). Therefore, in localities where rainbow trout establish self-sustaining populations, competition between rainbow trout and Atlantic salmon for habitats and food is possible.

Pink salmon fry migrate to sea in early summer, shortly after emerging from the gravel. Due to their rapid exodus from streams at emergence, pink salmon fry feed less in freshwater than other Pacific salmon. Hence, any competition for food between pink salmon and Atlantic salmon may take place during a short period in early summer, only.

In North America, interactions between brown trout and Atlantic salmon are thought to be highest during the first year of life when density-dependent processes are most intense (Milner et al. 2003).

In Russia, the White Sea rivers have two distinct runs of Atlantic salmon. The summer run salmon ascend the river in June-July and spawn in autumn the same year. Autumn run fish start their migration in early August and continue entering the river until it freezes. They do not spawn in the year they arrive. Autumn run salmon overwinter and stay in the rivers until they spawn in autumn the following year. Summer running fish are less numerous than autumn run fish. Pink salmon also enter the White Sea rivers in July and spawn in August whereas Atlantic salmon spawn in September and October. Typically pink salmon prefer shallower areas and do not compete with Atlantic salmon for territory in big rivers but this competition can occur in small rivers and in tributaries of big river systems when pink salmon enter streams in large numbers and aggressively push overwintered autumn run Atlantic salmon out of holding pools to non-typical habitats (Zubchenko et al. 2004).

In Canada, displacement, decreases in abundance and local extinction of other species has been observed following introduction of rainbow trout in a new ecosystem.

## Predation

Rainbow trout are an effective predator on fish, and several studies have demonstrated that rainbow trout have impacted local fish populations (Crowl et al., 1992; Behnke, 2002; Fausch, 2008). Hence, it is possible that rainbow trout may feed on Atlantic salmon eggs, fry and parr when they are present.

Adult pink salmon do not feed after entering freshwater (Heard, 1991), and hence, predation on Atlantic salmon fry and parr is not expected to occur.

## Hybridization

Hybridization rates between Atlantic salmon and brown trout are higher where one of the species is exotic than where both are native (Verspoor and Hammar, 1991; Allendorf et al., 2001). Factors influencing hybridization between Atlantic salmon and brown trout are poorly understood. However, hybrids are known to be viable (Day, 1884; Nygren et al., 1975; Hindar et al., 1997; Castillo et al., 2007). In North America, hybridization generally involves brown trout females (McGown and Davidson, 1992; Gephard et al., 2000) and mature male Atlantic salmon parr (Gephard et al., 2000; GarciaVazquez et al., 2001).

Hybridization between Atlantic salmon and rainbow trout is unlikely given that the species are from distinct genera and often have discrete spawning seasons.

The following table provides a general summary of potential threats to Atlantic salmon, and the relative likelihood of risk, from the presence of rainbow trout, pink salmon and brown trout where these occur as exotics:

| Potential threat | From rainbow trout <br> (outside their native <br> Rance) | From pink salmon <br> (outside their native <br> RANGE) | From brown trout <br> (outside their native <br> RanGe) |
| :--- | :--- | :--- | :--- |
| Spread of parasites | Very likely | Not evidenced | Not evidenced |
| Spread of diseases | Likely | Not evidenced | Not evidenced |
| Destruction of redds | Evidenced | Unlikely | Unlikley |
| Competition for resources <br> and areas | Likely | Likely, but for short <br> periods | Likely |
| Predation | Likely | Unlikely | Likely |
| Hydridisation | Unlikely | Unlikely | Unlikely |

### 2.6 Reports from ICES expert group reports relevant to North Atlantic salmon

### 2.6.1 WGRECORDS

The Working Group on the Science Requirements to Support Conservation, Restoration and Management of Diadromous Species (WGRECORDS) was established to provide a scientific forum in ICES for the coordination of work on diadromous species. The role of the Group is to coordinate work on diadromous species, organize Expert Groups, Theme Sessions and Symposia, and help to deliver the ICES Science Plan.

WGRECORDS held an informal meeting on 6th June 2012, during the NASCO Annual Meeting in Edinburgh, UK Scotland. Discussions were held on the requirements for Expert Groups to address new and ongoing issues on Atlantic salmon including issues arising from the NASCO Annual Meeting. The annual meeting of WGRECORDS was held on the 19th and 20th September 2012, during the ICES Annual Science Conference in Bergen, Norway. The meetings were chaired by Niall Ó Maoiléidigh (Ireland) and Atso Romakkaniemi (Finland) and attended by ten participants from eight countries.

The WGRECORDS Annual Meeting received reports from all the ICES Expert Groups working on diadromous species, and considered their progress and future requirements. Updates were received from three four expert groups of particular relevance to North Atlantic salmon which had been established by ICES following proposals by WGRECORDS. Summaries of all these expert groups are provided in this section.

- Workshop on Age Determination in Salmon (WKADS)-Section 2.6.2;
- Workshop on Data Collection Framework for Eels and Salmon (WKDCEF)Section 2.7.2;
- Workshop on Historical Tag Recovery Data (WKSTAR)-Section 2.6.3;
- Working Group on the Effectiveness of Recovery Actions for Atlantic Salmon (WGERAAS)-Section 2.4.

Other issues arising from the WGRECORDS meeting which are of particular relevance to Atlantic salmon were:

- Inclusion of new proposals for Atlantic salmon data collection under the EU DC-MAP.
- Proposals for future theme sessions at the ICES ASC e.g.:
- Implication of climate change for diadromous and migratory species over broad geographic scales;
- Parasites and diseases in a changing environment;
- Drug resistance in fish parasites and diseases;
- Changes in distribution of fish in response to climate change;
- Long-term planning to respond to effects of climate change on diadromous fish stocks;
- Climate change processes and predictions of impacts on salmon and eels;
- Carrying capacity and ecosystem interactions associated with mariculture.


### 2.6.2 WKADS 2

A second Workshop on Age Determination of Salmon (WKADS 2) took place from September 4th to 6th, 2012 in Derry, Londonderry, UK (Northern Ireland). Attended by 12 people from six countries, representing nine laboratories, the meeting addressed recommendations made at the previous WKADS meeting (ICES, 2011a) to review, assess, document and make recommendations for ageing and growth estimations of Atlantic salmon using digital scale reading, with a view to standardization. Available tools for measurement, quality control and implementation of inter-laboratory QC were considered.

Information on scale reading errors and inaccuracies was presented, including:

- possible scale deformation from jewellers press;
- differences in circuli number and spacings, on scales from different locations on smolts;
- measurements of smolt and adult scales made by different scale readers;
- measurements of adult scales made by the same scale reader.

The image collection gathered during WKADS was augmented by addition of scale images showing complexities in their growth, including scales with growth checks and repeat spawners. Available material detailing scale preparation, reading (microfiche, microscope and digital reading) and storage was reviewed and itemized, detailing the best practice pertinent to Atlantic salmon in one place.

Recommendations arising from the workshop were endorsed by WGNAS i.e.:

- An inter-lab calibration exercise should be held remotely in the next two to four years.
- Reference scale images and accompanying details should be hosted on ICES age readers forum website supported by the ICES PGCCDBS.
- The importance of the initial positioning of the line on a scale along which measurement are made, should be emphasized to all readers.

The Working Group recognize that there is still scope for further work comparing scale readings from different locations on salmon.

### 2.6.3 WKSTAR

The Workshop on Salmon Tagging Archive (WKSTAR) worked by correspondence in 2010/2011 and met at ICES Headquarters, Copenhagen, Denmark, 19-21 June 2012. The purpose of the Workshop was to ensure that the data compiled by the previous Workshops (ICES, 2007a; ICES, 2008b and 2009b) was fully archived and documented. These reports have resulted in recent peer reviewed publications in the ICES Journal of Marine Science in November 2012 (Reddin et al., 2012; Jacobsen et al., 2012) and presentation of results at the NASCO/ICES Salmon Symposium held in La Rochelle, France in October 2011. A resolution to ICES to record a summary of the workshops, presentations and publications in a Cooperative Research Report was accepted by ICES and the Workshop has developed an outline of the CRR.

The Workshop also undertook further checking and tidying of the tag recovery databases for Faroes and Greenland to identify and correct and resolve various anomalies in the datasets. Both Greenland and Faroes databases were updated, but it was noted that there were some gaps in the database. It was agreed that key scientists should be asked to consult with the "data owners" in the actual countries to give permission to include the data in the ICES database. This was organized by the chair of the WKSTAR and all data holders granted permission to develop the database.

The Working Group noted that the contact list would require updating upon completion of the database.

### 2.7 NASCO has asked ICES to provide a compilation of tag releases by country in 2012

### 2.7.1 Compilation of tag releases and fin clip data by ICES Member Countries in 2012

Data on releases of tagged, finclipped and otherwise marked salmon in 2012 were provided to the Working Group and are compiled as a separate report (ICES, 2013). In summary (Table 2.7.1.1), about 3.69 million salmon were marked in 2012, a decrease from the 4,18 million fish marked in 2011. The adipose clip was the most commonly used primary mark ( 3.145 million), with coded wire microtags ( 0.486 million) the next most common primary mark and 45582 fish were marked with external tags. Most marks were applied to hatchery origin juveniles ( 3.620 million), while 58465 wild juveniles and 6800 adults were also marked. In 2012, 15126 PIT tagged salmon, Data Storage Tags (DSTs), radio and/or sonic transmitting tags (pingers) were also used (Table 2.7.1.1).

From 2003, the Working Group has recorded information on marks being applied to farmed salmon. These may help trace the origin of farmed salmon captured in the wild in the case of escape events. Two jurisdictions (USA and Iceland) have required that some or all of the sea cage farmed fish reared in their area be marked. In Iceland, coded wire tags have been applied to about $5-10 \%$ of sea cage farm production in certain areas. The use of genetic marked salmon will from 2012 gradually replace microtagged fish in aquaculture in Iceland. In USA, the industry has opted for a genetic "marking" procedure. The broodstock has been screened with molecular genetic techniques, which makes it feasible to trace an escaped farmed salmon back to its hatchery of origin and its farm site
through analysis of its DNA. Genetic assignment has also been applied for hatchery juveniles that are released in two large rivers in the Southwest of France.

### 2.8 NASCO has asked ICES to identify relevant data deficiencies, monitoring needs and research requirements

### 2.8.1 NASCO subgroup on marine research

NASCO has asked ICES to identify relevant data deficiencies, monitoring needs and research requirements. In considering this question, the Working Group discussed the report of the NASCO Subgroup on the Future Direction of Research on Marine Survival of Salmon which met in London in December, 2012. This subgroup was convened by the International Atlantic Salmon Research Board to evaluate recent scientific progress in studies of marine mortality of salmon and to provide guidance on how the Board's Scientific Advisory Group can remain an effective and productive group in future.

The subgroup reviewed the findings of recent scientific investigations into the causes of increased mortality at sea and the implications of these findings for management. It noted that genetic stock identification and other advances in the field of genetics, migration modelling, tracking and studies of the diet of salmon at sea developed under the SALSEA programme all have considerable implications for salmon management. The subgroup also reviewed the Board's inventory of research and identified opportunities for enhanced collaboration, gaps in the research programme and future research needs to support management. It considered that analysing the remaining samples and data arising from the SALSEA-Merge, SALSEA West Greenland and SALSEA North America programmes should be a priority.

The subgroup also proposed that a particular focus for the IASRB should be studies to partition marine mortality of salmon among the phases of the marine migration, and it recommended that the IASRB should consider facilitating a meeting of scientists and external partners to further develop a collaborative international programme of research in this area. The subgroup also developed an outline proposal for acoustically tagging emigrating smolts and tracking their movements with detector arrays, and other novel detector systems, noting that analytical techniques were now being applied to such data collected in North America to partition the mortality during the early stages of the marine phase.

The Working Group considered these recommendations alongside their own evaluation of current research needs. They endorsed the view of the subgroup that analysis of outstanding samples during the marine surveys under the SALSEA programme should be a priority and that mechanisms should be sought to obtain funding to support this.

The Working Group reviewed the proposal outlining a collaborative international programme of research on marine mortality of salmon provided by the subgroup. The outline described a project to estimate stage-specific mortality rates of marine salmon by using acoustic technologies to monitor migrating Atlantic salmon. The project would build on the existing infrastructure and historical datasets from index rivers in NAC and NEAC areas, would apply knowledge gained from SALSEA activities on timing and migration corridors of post-smolts in southern NEAC and from advances in acoustic tracking technologies (Whoriskey, 2011; Lacroix, 2013), and would benefit from academ-
ic, industry and government partnerships. Emigrating smolts released from specific index rivers throughout the NAC and NEAC regions would be tagged with acoustic tags and tagged smolts would be tracked throughout the river, estuary and marine environments via strategically placed ultrasonic telemetry receiver arrays at identified choke points along the nearshore migration paths of post-smolts and at locations associated with other marine research and monitoring activities (e.g. buoy deployments for oceanographic monitoring, research survey cruises, wave gliders, etc.). Estimates of survival probabilities could be obtained by applying a variety of statistical methods and models to the resulting data (see Section 2.3.4.3).

The Working Group endorsed the proposed project outline. It was noted that this type of acoustic monitoring of marine phase salmon is currently underway in NAC. Large numbers of smolts are being tagged and their migration is being monitored via ultrasonic receiver arrays hundreds of kilometres into the marine environment in Canada (see Section 2.3.4.2) and US (ICES, 2012a).

The Working Group recommends that the IASRB support the further development of the project outlined by the NASCO Subgroup on the Future Direction of Research on Marine Survival of Salmon. A large international coordinated project monitoring the marine migration of many salmon stocks across the North Atlantic may provide stage-specific estimates of marine survival that would increase knowledge of marine ecology and better inform management. Stage-specific marine mortality estimates would help improve essential inputs in stock assessment models and would provide additional information for testing hypotheses on the causal mechanisms for the increase in marine mortality documented for most stocks across the North Atlantic in recent decades. These results would also be of benefits for managers trying to identify areas where action might be taken to mitigate current impacts. Detailed information on migration dynamics of salmon in nearshore waters will also aid managers involved in marine spatial planning to evaluate the impacts of alternative/renewable energy projects (e.g. wind energy, tidal energy, etc.) in marine waters.

The Working Group encourages the IASRB to consider expanding the focus of this research project beyond the scope of salmon. Integrating the research needs across different species would increase the benefit of an effort like this and increase the likelihood of successfully competing for funding support. The Working Group also encourages the IASRB to consider the wide variety of resources and experiences available for an endeavour such as this. Large-scale multinational tracking programmes are already underway in NAC and the experience gained from these efforts would increase the likelihood of success for any effort initiated in NEAC. It was noted that many tracking projects have previously been conducted in Norway and UK (Scotland) (for examples see Middlemas et al., 2009; Thorstad et al., 2012a; Thorstad et al., 2012b; Davidsen et al., 2013) although a large international collaborative effort has not been conducted to date.

The Working Group recognizes that consideration for ultrasonically tagging and releasing non-maturing salmon captured at Greenland. A significant ultrasonic array exists with the NAC area. Considering that the North American contribution to the Greenland harvest has averaged $80 \%$ since 2003 (Table 5.1.2.1), there is a high likelihood that any tagged salmon would be of NAC origin (with the potential for determining river of origin via genetic analysis) and may be detected during their return migration to their
natal river depending on where they are migrating to and the mortality experienced from tagging to homewater. Tagging effort could be combined with future sampling satellite tagging efforts if undertaken (see Section 2.3.4.1). Information on survival during the second winter at sea may help improve essential inputs in stock assessment models and would provide additional information for testing hypotheses on the causal mechanisms for the increase in marine mortality documented for most stocks across the North Atlantic in recent decades.

The Working Group noted that the NASCO subgroup had advised that the SAG is the only body within NASCO that identifies research needs and addresses scientific coordination. It concluded that it is the most appropriate and effective forum in which to perform this important role. The Working Group endorsed this view, noting that the SAG provided an essential mechanism for scientists to collaboratively work with managers to develop scientific programmes to support the conservation, protection and enhancement of salmon stocks.

### 2.8.2 Workshop on Eel and Salmon Data Collection Framework (WKESDCF)

PGCCDS: The Workshop on Eel and Salmon Data Collection Framework met in Copenhagen in July 2012, under the co-chairmanship of Ted Potter (UK) and Alan Walker (UK) and was attended by 23 experts in eel and salmon assessment and management, representing nine EU Member States. Changes to the EU Data Collection Framework Data Collection - Multi-Annual Programme (DCF) in 2007 introduced requirements to collect data on eel and salmon, but the specific data requested for these species did not meet the needs of national and international assessments. The EC (DGMare) has indicated that they intend to simplify the rules and formats within DC-MAP and increase the flexibility for data collection programmes. Thus many of the details of the data collection programmes will be agreed by Regional Coordination Groups (RCG). There will be also greater focus on the needs of end-users (e.g. ICES) who will be asked to provide feedback on the quality of data provided for assessment purposes. The proposed development of the new Data Collection - Multi-Annual Programme (DC-MAP) in 2013-2014 provides the opportunity to coordinate and improve the collection of data used in assessments for these species.

The key tasks of the workshop were to:

- Determine the data required to support international obligations for the assessment of eel and salmon;
- Describe the national monitoring and survey programmes required to meet these data requirements; and
- Consider options for integrating salmon and eel surveys and monitoring.

For each species/area, the Workshop considered: the national/international management objectives; the assessments undertaken to support these objectives; the data required to undertake the assessments; and proposed changes to the DC-MAP to provide these data. The existing DCF also requires the collection of data on economics and aquaculture; these data are important in the management of diadromous species, but the workshop did not contain the expertise necessary to consider these elements in detail.

Eel and salmon differ markedly from marine species in their biology, the nature and distribution of their fisheries, and the methods used to assess stock status and provide management advice. As a result, the data collection requirements do not fit well into the 'standard' approaches used for marine species. In particular, much of the assessment of both species is conducted at a local and national level even when the results contribute to international assessments (e.g. development of Conservation Limits for salmon river stocks). These approaches may differ depending upon a range of factors including the practicalities of collecting particular data.

The Workshop made detailed recommendations for several tiers of data collection. Some data (e.g. catches) are required for all stock components and are of little value if they are not collected in a consistent way for all stocks/fisheries. The collection of other data may depend on local requirements and constraints, for example to support the local development of river-specific conservation limits. The Working Group endorsed the proposals for data collection on Atlantic salmon proposed by the workshop. The workshop report has also been considered by the EU Scientific, Technical and Economic Committee for Fisheries (STECF) as part of the review of the DC-MAP proposals. STECF endorsed the recommendation that DC-MAP should include the requirement to collect salmon data needed for stock assessment purposes and that, if possible, this should include data collected in inland waters including from recreational fisheries. However, they noted that the WKESDFC recommendations were too detailed to be included in the DC-MAP in full because the intention was to keep the DC-MAP simple and flexible. STECF therefore proposed that the details of the data collection for salmon, including the choice of index rivers and variables, should be agreed by appropriate RCGs. The Working Group were concerned that if these decisions were made by different RCGs for different regions, it would inevitably result in differences in the data collection procedures which may cause problems for subsequent assessment work. The Working Group therefore recommended the establishment of an RCG for diadromous species to consider the unified collection of data on all salmon stocks (as well as eel).

DGMare has also provided feedback on the workshop report, indicating that they found Table 4.2.3.1 of the WKESDCF report, which provides an overview of the compatibility of data collected under the DCF with the data needed for the assessment of Baltic salmon by ICES, particularly helpful. Following a request from ICES a table containing an overview of the compatibility of data currently collected under the DCF with the data available, reviewed, and needed for the annual assessment of North Atlantic salmon by ICES was compiled (Table 2.8.2.1).

The Workshop also identified a number of areas where coordinated data collection might offer opportunities for increased cost-effectiveness in some circumstances, including: electric fishing surveys; trapping programmes; operation of automatic counters; and habitat surveys.

### 2.8.3 Stock annex development

The Working Group considered proposals from the Review Group regarding the establishment of an Atlantic salmon stock annex. Such stock annexes have been developed for other ICES assessment WG reports and are intended to provide a complete description of the methodology used in conducting stock assessments and the provision of catch advice. These documents are intended to be informative for members of the WG and reviewers
as well as aid communication with the general public. The Working Group noted that the Baltic Salmon Working Group (WGBAST) had developed such an annex as part of their recent Inter-Benchmark Protocol exercise (ICES, 2012b).

The Working Group agreed that the development of a specific Atlantic salmon stock annex would be helpful and to take forward this initiative. Initial progress was made in completing a first draft, largely by compiling information contained in earlier WGNAS reports and other sources. However, the Working Group had insufficient time to complete the task during the 2013 meeting. It further noted that aspects of the annex would require country-specific inputs. To develop these, the Working Group recommended that an Atlantic salmon stock annex should be developed using an agreed template and that country specific inputs should be prepared for the 2014 meeting with a view to finalizing the document at that time.

Table 2.1.1.1. Reported total nominal catch of salmon by country (in tonnes round fresh weight), 1960-2012. (2012 figures include provisional data).

| Year | NAC Area |  |  | NEAC (N. Area) |  |  |  |  |  |  | NEAC (S. Area) |  |  |  |  |  | Faroes \& Greenland |  |  |  | Total <br> Reported <br> Nominal <br> Catch | Unreported catches |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Canada <br> (1) |  | St. P\&M | Norway <br> (2) | Russia <br> (3) | $\begin{gathered} \text { Icel } \\ \hline \text { Wild } \end{gathered}$ | $\frac{\text { land }}{\text { Ranch (4) }}$ | Sweden <br> (West) | Denmark | Finland | Ireland $(5,6)$ | $\begin{gathered} \text { UK } \\ (\mathrm{E} \& \mathrm{~W}) \end{gathered}$ |  | $\begin{gathered} \text { UK } \\ (\mathrm{Scotl} .) \end{gathered}$ | France (8) | Spain (9) | $\begin{gathered} \text { Faroes } \\ (10) \end{gathered}$ | East Grld. | West <br> Grld. <br> (11) | $\begin{gathered} \text { Other } \\ (12) \end{gathered}$ |  | $\begin{gathered} \text { NASCO } \\ \text { Areas (13) } \end{gathered}$ | International <br> waters (14) |
| 1960 | 1,636 | 1 | - | 1,659 | 1,100 | 100 | - | 40 | - | - | 743 | 283 | 139 | 1,443 | - | 33 | - | - | 60 | - | 7,237 | - | - |
| 1961 | 1,583 | 1 | - | 1,533 | 790 | 127 | - | 27 | - | - | 707 | 232 | 132 | 1,185 | - | 20 | - | - | 127 | - | 6,464 | - | - |
| 1962 | 1,719 | 1 | - | 1,935 | 710 | 125 | - | 45 | - | - | 1,459 | 318 | 356 | 1,738 | - | 23 | - | - | 244 | - | 8,673 | - | - |
| 1963 | 1,861 | 1 | - | 1,786 | 480 | 145 | - | 23 | - | - | 1,458 | 325 | 306 | 1,725 | - | 28 | - | - | 466 | - | 8,604 | - | - |
| 1964 | 2,069 | 1 | - | 2,147 | 590 | 135 | - | 36 | - | - | 1,617 | 307 | 377 | 1,907 | - | 34 | - | - | 1,539 | - | 10,759 | - | - |
| 1965 | 2,116 | 1 | - | 2,000 | 590 | 133 | - | 40 | - | - | 1,457 | 320 | 281 | 1,593 | - | 42 | - | - | 861 | - | 9,434 | - | - |
| 1966 | 2,369 | 1 | - | 1,791 | 570 | 104 | 2 | 36 | - | - | 1,238 | 387 | 287 | 1,595 | - | 42 | - | - | 1,370 | - | 9,792 | - | - |
| 1967 | 2,863 | 1 | - | 1,980 | 883 | 144 | 2 | 25 | - | - | 1,463 | 420 | 449 | 2,117 | - | 43 | - | - | 1,601 | - | 11,991 | - | - |
| 1968 | 2,111 | 1 | - | 1,514 | 827 | 161 | 1 | 20 | - | - | 1,413 | 282 | 312 | 1,578 | - | 38 | 5 | - | 1,127 | 403 | 9,793 | - | - |
| 1969 | 2,202 | 1 | - | 1,383 | 360 | 131 | 2 | 22 | - | - | 1,730 | 377 | 267 | 1,955 | - | 54 | 7 | - | 2,210 | 893 | 11,594 | - | - |
| 1970 | 2,323 | 1 | - | 1,171 | 448 | 182 | 13 | 20 | - | - | 1,787 | 527 | 297 | 1,392 | - | 45 | 12 | - | 2,146 | 922 | 11,286 | - | - |
| 1971 | 1,992 | 1 | - | 1,207 | 417 | 196 | 8 | 18 | - | - | 1,639 | 426 | 234 | 1,421 | - | 16 | - | - | 2,689 | 471 | 10,735 | - | - |
| 1972 | 1,759 | 1 | - | 1,578 | 462 | 245 | 5 | 18 | - | 32 | 1,804 | 442 | 210 | 1,727 | 34 | 40 | 9 | - | 2,113 | 486 | 10,965 | - | - |
| 1973 | 2,434 | 3 | - | 1,726 | 772 | 148 | 8 | 23 | - | 50 | 1,930 | 450 | 182 | 2,006 | 12 | 24 | 28 | - | 2,341 | 533 | 12,670 | - | - |
| 1974 | 2,539 | 1 | - | 1,633 | 709 | 215 | 10 | 32 | - | 76 | 2,128 | 383 | 184 | 1,628 | 13 | 16 | 20 | - | 1,917 | 373 | 11,877 | - | - |
| 1975 | 2,485 | 2 | - | 1,537 | 811 | 145 | 21 | 26 | - | 76 | 2,216 | 447 | 164 | 1,621 | 25 | 27 | 28 | - | 2,030 | 475 | 12,136 | - | - |
| 1976 | 2,506 | 1 | 3 | 1,530 | 542 | 216 | 9 | 20 | - | 66 | 1,561 | 208 | 113 | 1,019 | 9 | 21 | 40 | $<1$ | 1,175 | 289 | 9,327 | - | - |
| 1977 | 2,545 | 2 | - | 1,488 | 497 | 123 | 7 | 10 | - | 59 | 1,372 | 345 | 110 | 1,160 | 19 | 19 | 40 | 6 | 1,420 | 192 | 9,414 | - | - |
| 1978 | 1,545 | 4 | - | 1,050 | 476 | 285 | 6 | 10 | - | 37 | 1,230 | 349 | 148 | 1,323 | 20 | 32 | 37 | 8 | 984 | 138 | 7,682 | - | - |
| 1979 | 1,287 | 3 | - | 1,831 | 455 | 219 | 6 | 12 | - | 26 | 1,097 | 261 | 99 | 1,076 | 10 | 29 | 119 | $<0,5$ | 1,395 | 193 | 8,118 | - | - |
| 1980 | 2,680 | 6 | - | 1,830 | 664 | 241 | 8 | 17 | - | 34 | 947 | 360 | 122 | 1,134 | 30 | 47 | 536 | $<0,5$ | 1,194 | 277 | 10,127 | - | - |
| 1981 | 2,437 | 6 | - | 1,656 | 463 | 147 | 16 | 26 | - | 44 | 685 | 493 | 101 | 1,233 | 20 | 25 | 1,025 | $<0,5$ | 1,264 | 313 | 9,954 | - | - |
| 1982 | 1,798 | 6 | - | 1,348 | 364 | 130 | 17 | 25 | - | 54 | 993 | 286 | 132 | 1,092 | 20 | 10 | 606 | $<0,5$ | 1,077 | 437 | 8,395 | - | - |
| 1983 | 1,424 | 1 | 3 | 1,550 | 507 | 166 | 32 | 28 | - | 58 | 1,656 | 429 | 187 | 1,221 | 16 | 23 | 678 | $<0,5$ | 310 | 466 | 8,755 | - | - |
| 1984 | 1,112 | 2 | 3 | 1,623 | 593 | 139 | 20 | 40 | - | 46 | 829 | 345 | 78 | 1,013 | 25 | 18 | 628 | $<0,5$ | 297 | 101 | 6,912 | - | - |
| 1985 | 1,133 | 2 | 3 | 1,561 | 659 | 162 | 55 | 45 | - | 49 | 1,595 | 361 | 98 | 913 | 22 | 13 | 566 | 7 | 864 | - | 8,108 | - | - |
| 1986 | 1,559 | 2 | 3 | 1,598 | 608 | 232 | 59 | 54 | - | 37 | 1,730 | 430 | 109 | 1,271 | 28 | 27 | 530 | 19 | 960 | - | 9,255 | 315 | - |
| 1987 | 1,784 | 1 | 2 | 1,385 | 564 | 181 | 40 | 47 | - | 49 | 1,239 | 302 | 56 | 922 | 27 | 18 | 576 | $<0,5$ | 966 | - | 8,159 | 2,788 | - |
| 1988 | 1,310 | 1 | 2 | 1,076 | 420 | 217 | 180 | 40 | - | 36 | 1,874 | 395 | 114 | 882 | 32 | 18 | 243 | 4 | 893 | - | 7,737 | 3,248 | - |
| 1989 | 1,139 | 2 | 2 | 905 | 364 | 141 | 136 | 29 | - | 52 | 1,079 | 296 | 142 | 895 | 14 | 7 | 364 | - | 337 | - | 5,904 | 2,277 | - |
| 1990 | 911 | 2 | 2 | 930 | 313 | 141 | 285 | 33 | 13 | 60 | 567 | 338 | 94 | 624 | 15 | 7 | 315 | - | 274 | - | 4,925 | 1,890 | 180-350 |

Table 2.1.1.1. Continued.

| Year | NAC Area |  |  | NEAC (N. Area) |  |  |  |  |  |  | NEAC (S. Area) |  |  |  |  |  | Faroes \& Greenland |  |  |  | Total Reported Nominal Catch | Unreported catches |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Canada <br> (1) | USA | St. P\&M | Nonway <br> (2) | Russia <br> (3) |  | $\begin{aligned} & \text { eland } \\ & \text { Ranch (4) } \\ & \hline \end{aligned}$ | Sweden <br> (West) | Denmark | Finland | Ireland $(5,6)$ | $\begin{gathered} \text { UK } \\ (\mathrm{E} \& \mathrm{~W}) \end{gathered}$ | $\begin{gathered} \text { UK } \\ (\text { (N.IrL.) } \\ (6,7) \end{gathered}$ | $\begin{gathered} \text { UK } \\ (\text { Scotl. }) \end{gathered}$ | France <br> (8) | $\begin{gathered} \text { Spain } \\ (9) \\ \hline \end{gathered}$ | $\begin{gathered} \text { Faroes } \\ (10) \\ \hline \end{gathered}$ | East Grld. |  | $\begin{gathered} \text { Other } \\ (12) \\ \hline \end{gathered}$ |  | $\begin{gathered} \text { NASCO } \\ \text { Areas (13) } \end{gathered}$ | International <br> waters (14) |
| 1991 | 711 | 1 | 1 | 876 | 215 | 129 | 346 | 38 | 3 | 70 | 404 | 200 | 55 | 462 | 13 | 11 | 95 | 4 | 472 | - | 4,106 | 1,682 | 25-100 |
| 1992 | 522 | 1 | 2 | 867 | 167 | 174 | 462 | 49 | 10 | 77 | 630 | 171 | 91 | 600 | 20 | 11 | 23 | 5 | 237 | - | 4,119 | 1,962 | 25-100 |
| 1993 | 373 | 1 | 3 | 923 | 139 | 157 | 499 | 56 | 9 | 70 | 541 | 248 | 83 | 547 | 16 | 8 | 23 | - | - | - | 3,696 | 1,644 | 25-100 |
| 1994 | 355 | 0 | 3 | 996 | 141 | 136 | 313 | 44 | 6 | 49 | 804 | 324 | 91 | 649 | 18 | 10 | 6 | - | - | - | 3,945 | 1,276 | 25-100 |
| 1995 | 260 | 0 | 1 | 839 | 128 | 146 | 303 | 37 | 3 | 48 | 790 | 295 | 83 | 588 | 10 | 9 | 5 | 2 | 83 | - | 3,629 | 1,060 | - |
| 1996 | 292 | 0 | 2 | 787 | 131 | 118 | 243 | 33 | 2 | 44 | 685 | 183 | 77 | 427 | 13 | 7 | - | 0 | 92 | - | 3,136 | 1,123 | - |
| 1997 | 229 | 0 | 2 | 630 | 111 | 97 | 59 | 19 | 1 | 45 | 570 | 142 | 93 | 296 | 8 | 4 | $\cdot$ | 1 | 58 | - | 2,364 | 827 | - |
| 1998 | 157 | 0 | 2 | 740 | 131 | 119 | 46 | 15 | 1 | 48 | 624 | 123 | 78 | 283 | 8 | 4 | 6 | 0 | 11 | - | 2,395 | 1,210 | - |
| 1999 | 152 | 0 | 2 | 811 | 103 | 111 | 35 | 16 | 1 | 62 | 515 | 150 | 53 | 199 | 11 | 6 | 0 | 0 | 19 | - | 2,247 | 1,032 | - |
| 2000 | 153 | 0 | 2 | 1,176 | 124 | 73 | 11 | 33 | 5 | 95 | 621 | 219 | 78 | 274 | 11 | 7 | 8 | 0 | 21 | - | 2,912 | 1,269 | - |
| 2001 | 148 | 0 | 2 | 1,267 | 114 | 74 | 14 | 33 | 6 | 126 | 730 | 184 | 53 | 251 | 11 | 13 | 0 | 0 | 43 | - | 3,069 | 1,180 | - |
| 2002 | 148 | 0 | 2 | 1,019 | 118 | 90 | 7 | 28 | 5 | 93 | 682 | 161 | 81 | 191 | 11 | 9 | 0 | 0 | 9 | - | 2,654 | 1,039 | - |
| 2003 | 141 | 0 | 3 | 1.071 | 107 | 99 | 11 | 25 | 4 | 78 | 551 | 89 | 56 | 192 | 13 | 9 | 0 | 0 | 9 | - | 2.457 | 847 | - |
| 2004 | 161 | 0 | 3 | 784 | 82 | 111 | 18 | 20 | 4 | 39 | 489 | 111 | 48 | 245 | 19 | 7 | 0 | 0 | 15 | - | 2.157 | 686 | - |
| 2005 | 139 | 0 | 3 | 888 | 82 | 129 | 21 | 15 | 8 | 47 | 422 | 97 | 52 | 215 | 11 | 13 | 0 | 0 | 15 | - | 2.156 | 700 | - |
| 2006 | 137 | 0 | 3 | 932 | 91 | 93 | 17 | 14 | 2 | 67 | 326 | 80 | 29 | 192 | 13 | 11 | 0 | 0 | 22 | - | 2.029 | 670 | , |
| 2007 | 112 | 0 | 2 | 767 | 63 | 93 | 36 | 16 | 3 | 58 | 85 | 67 | 30 | 171 | 11 | 9 | 0 | 0 | 25 | - | 1.548 | 475 | - |
| 2008 | 158 | 0 | 4 | 807 | 73 | 132 | 69 | 18 | 9 | 71 | 89 | 64 | 21 | 161 | 12 | 9 | 0 | 0 | 26 | - | 1,721 | 443 | - |
| 2009 | 126 | 0 | 3 | 595 | 71 | 126 | 44 | 17 | 8 | 36 | 68 | 54 | 17 | 121 | 4 | 2 | 0 | 0 | 26 | - | 1,318 | 327 | - |
| 2010 | 153 | 0 | 3 | 642 | 88 | 147 | 42 | 22 | 13 | 49 | 99 | 109 | 12 | 180 | 10 | 2 | 0 | 0 | 40 | - | 1,610 | 367 | - |
| 2011 | 179 | 0 | 4 | 696 | 89 | 98 | 30 | 39 | 13 | 44 | 87 | 136 | 10 | 159 | 11 | 7 | 0 | 0 | 28 | - | 1,629 | 421 | - |
| 2012 | 135 | 0 | 1 | 696 | 82 | 53 | 12 | 30 | 2 | 64 | 88 | 57 | 9 | 130 | 10 | 8 | 0 | 0 | 33 | - | 1,409 | 403 | $\cdot$ |
| Average $2007-2011$ 2002 -2011 | 146 145 | 0 0 | 3 3 | 701 820 | 77 86 | 119 112 | 44 29 | 22 21 | 9 7 | 52 58 | $\begin{gathered} 85 \\ 290 \end{gathered}$ | $\begin{aligned} & 86 \\ & 97 \end{aligned}$ | 18 36 | 158 183 | 9 11 | 6 8 | 0 | 0 0 | 29 22 | . | 1,565 1,928 | 407 598 | . |
| 2002-2011 | 145 | 0 | 3 | 820 | 86 | 112 | 29 | 21 | 7 | 58 | $290$ | $97$ | 36 | 183 | 11 | 8 | 0 | 0 | 22 | - | 1.928 | 598 | $\cdot$ |

1. Includes estimates of some local sales, and, prior to 1984 , by-catch
2. Before 1966 , sea trout and sea charr included ( $5 \%$ of total)
3. Figures from 1991 to 2000 do not include catches taken in the recreational (rod) fishery.
4 From 1990, catch includes fish ranched for both commercial and angling purposes.
4. Improved reporting of rod catches in 1994 and data derived from carcase tagging and log books from 2002
5. Catch on River Foyle allocated $50 \%$ Ireland and $50 \% \mathrm{~N}$. Ireland.
6. Angling catch (derived from carcase tageing and log books) first included in 2002.
7. Data for France include some urreported catches
. Weights estimated from mean weight of tish caught in Astunas ( $80.90 \%$ of Spanish catch)
8. Between $1991 \& 1999$, there was only a research fishery at Faroes. In $1997 \& 1999$ no fishery took place;
the conmer cial fishery resumed in 2000 , but has not operated since 2001
9. Includes catches made in the West Greenland area by Norway, Faroes.

Sweden and Denmark in 1965-1975.
12. Includes catches in Norwegian Sea by vessels from Demmakk, Sweden, Gemmany, Norway and Fiuland
13. No unreported catch estimate available for Canada in 2007 and 2008

Data for Canada in 2009 and 2010 are incomplete.
No ureported catch estimate available for Russia since 2008
14. Estimates refer to season ending in given year.

Table 2．1．1．2．Reported total nominal catch of salmon in home waters by country（in tonnes round fresh weight），1960－2012．（2012 figures include provisional data）． $\mathrm{S}=$ Salmon（ 2 SW or MSW fish）． $\mathrm{G}=\mathrm{Grilse}$（ 1 SW fish）． $\mathrm{Sm}=$ small． $\mathrm{Lg}=$ large； $\mathrm{T}=\mathrm{S}+\mathrm{G}$ or $\mathrm{Lg}+\mathrm{Sm}$ ．

| Year | NAC Area |  |  |  | NEAC（N．Area） |  |  |  |  |  |  |  |  |  |  | NEAC（S．Area） |  |  |  |  |  |  |  |  |  | $\begin{gathered} \text { Total } \\ \mathrm{T} \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Canada（1） |  |  | $\begin{gathered} \text { USA } \\ \text { T } \\ \hline \end{gathered}$ | Norway（2） |  |  | $\begin{gathered} \text { Russia } \\ \quad \begin{array}{c} (3) \\ T \\ \hline \end{array} ⿳ 亠 口 子 \end{gathered}$ | Iceland |  | Sweden （West） T | $\begin{gathered} \text { Deamark } \\ =T \\ = \end{gathered}$ | Finland |  |  | $\begin{aligned} & \text { Ireland } \\ & (4,5) \end{aligned}$ |  |  | $\begin{gathered} \text { UK } \\ (\text { E\&W }) \\ \mathbf{T} \end{gathered}$ | $\begin{gathered} \text { UK(N.L) } \\ \left(\begin{array}{c} 1,6) \\ T \end{array}\right. \\ \hline \end{gathered}$ | UK（Scotland） |  |  | $\begin{gathered} \text { France } \\ \mathrm{T} \end{gathered}$ | $\begin{gathered} \text { Spain } \\ \mathrm{T} \end{gathered}$ |  |
|  |  |  |  | Wid |  |  |  | Ranch |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | L2 | Sm | T |  | S | G | T |  | T | T |  |  | s | G | T | s | G | T |  |  | S | G | T |  |  |  |
| 1960 | － | － | 1，636 |  | 1 | － | － |  | 1，659 | 1，100 | 100 | － | 40 | － | － | － | － | － | － | 743 | 283 | 139 | 971 | 472 | 1，443 | － | 33 | 7，177 |
| 1961 | － | － | 1，583 | 1 | － | － | 1，533 | 790 | 127 | － | 27 | － | － | － | － | － | － | 707 | 232 | 132 | 811 | 374 | 1，185 | － | 20 | 6，337 |
| 1962 | － | － | 1，719 | 1 | － | － | 1,935 | 710 | 125 | － | 45 | － | － | － | － | － | － | 1.459 | 318 | 356 | 1，014 | 724 | 1，738 | － | 23 | 8，429 |
| 1963 | － | － | 1.861 | 1 | － | － | 1.786 | 480 | 145 | － | 23 | － | $\cdot$ | － | － | － | － | 1.458 | 325 | 306 | 1.308 | 417 | 1.725 | － | 28 | 8，138 |
| 1964 | － | － | 2.069 | 1 | － | － | 2.147 | 590 | 135 | － | 36 | － | － | － | － | － | － | 1.617 | 307 | 377 | 1.210 | 697 | 1.907 | － | 34 | 9，220 |
| 1965 | － | － | 2.116 | 1 | － | － | 2.000 | 590 | 133 | － | 40 | － | $\cdot$ | － | － | － | － | 1.457 | 320 | 281 | 1.043 | 550 | 1.593 | － | 42 | 8.573 |
| 1966 | － | － | 2.369 | 1 | － | $\cdot$ | 1.791 | 570 | 104 | 2 | 36 | － | $\cdot$ | － | － | － | － | 1.238 | 387 | 287 | 1.049 | 546 | 1.595 | － | 42 | 8.422 |
| 1967 | － | － | 2，863 | 1 | － | － | 1.980 | 883 | 144 | 2 | 25 | － | $\cdot$ | － | － | － | － | 1.463 | 420 | 449 | 1，233 | 88.4 | 2，117 | － | 43 | 10,390 |
| 1968 | － | － | 2，111 | 1 | － | － | 1.514 | 827 | 161 | 1 | 20 | － | － | － | － | － | － | 1，413 | 282 | 312 | 1，021 | 557 | 1，578 | － | 38 | 8，258 |
| 1969 | － | － | 2，202 | 1 | 801 | 582 | 1,383 | 360 | 131 | 2 | 22 | － | $\cdot$ | － | － | － | － | 1，730 | 377 | 267 | 997 | 958 | 1.955 | － | 54 | 8，484 |
| 1970 | 1，562 | 761 | 2,323 | 1 | 815 | 356 | 1，171 | 448 | 182 | 13 | 20 | － | $\cdot$ | － | － | － | － | 1，787 | 527 | 297 | 775 | 617 | 1，392 | － | 45 | 8，206 |
| 1971 | 1.482 | 510 | 1,992 | 1 | 771 | 436 | 1，207 | 417 | 196 | 8 | 18 | ． | $\cdot$ | － | ． | － | － | 1，639 | 426 | 234 | 719 | 702 | 1，421 | － | 16 | 7.575 |
| 1972 | 1，201 | 558 | 1，759 | 1 | 1，064 | 514 | 1,578 | 462 | 245 | 5 | 18 | － | － | － | 32 | 200 | 1，604 | 1，804 | 442 | 210 | 1，013 | 714 | 1，727 | 34 | 40 | 8，357 |
| 1973 | 1，651 | 783 | 2，434 | 3 | 1，220 | 506 | 1,726 | 772 | 148 | 8 | 23 | － | － | － | 50 | 244 | 1，686 | 1.930 | 450 | 182 | 1，158 | 848 | 2，006 | 12 | 24 | 9，768 |
| 1974 | 1，589 | 950 | 2，539 | 1 | 1，149 | 484 | 1，633 | 709 | 215 | 10 | 32 | － | $\cdot$ | － | 76 | 170 | 1.958 | 2.128 | 383 | 184 | 912 | 716 | 1，628 | 13 | 16 | 9，567 |
| 1975 | 1，573 | 912 | 2，485 | 2 | 1，038 | 499 | 1,537 | 811 | 145 | 21 | 26 | － | $\cdot$ | － | 76 | 274 | 1.942 | 2，216 | 447 | 164 | 1，007 | 614 | 1，621 | 25 | 27 | 9，603 |
| 1976 | 1，721 | 785 | 2，506 | 1 | 1，063 | 467 | 1.530 | 542 | 216 | 9 | 20 | － | $\cdot$ | － | ${ }_{6} 6$ | 109 | 1.452 | 1.561 | 208 | 113 | 522 | 497 | 1，019 | 9 | 21 | 7，821 |
| 1977 | 1，883 | 662 | 2，545 | 2 | 1，018 | 470 | 1，488 | 497 | 123 | 7 | 10 | － | － | － | 59 | 145 | 1，227 | 1，372 | 345 | 110 | 639 | 521 | 1，160 | 19 | 19 | 7，756 |
| 1978 | 1.225 | 320 | 1，545 | 4 | 668 | 382 | 1，050 | 476 | 285 | 6 | 10 | － | $\cdot$ | － | 37 | 147 | 1，082 | 1，229 | 349 | 148 | ${ }^{781}$ | 542 | 1，323 | 20 | 32 | 6，514 |
| 1979 | 705 | 582 | 1，287 | 3 | 1，150 | 681 | 1，831 | 455 | 219 | 6 | 12 | － | － | － | 26 | 105 | 922 | 1,027 | 261 | 99 | 598 | 478 | 1，076 | 10 | 29 | 6,341 |
| 1980 | 1，763 | 917 | 2，680 | 6 | 1,352 | 478 | 1，830 | 664 | 241 | 8 | 17 | － | ． | － | 34 | 202 | 745 | 947 | 360 | 122 | 851 | 283 | 1.134 | 30 | 47 | 8，120 |
| 1981 | 1，619 | 818 | 2，437 | 6 | 1，189 | 467 | 1.656 | 463 | 147 | 16 | 26 | － | － | － | 44 | 164 | 521 | 685 | 493 | 101 | 844 | 389 | 1，233 | 20 | 25 | 7，352 |
| 1982 | 1，082 | 716 | 1，798 | 6 | 985 | 363 | 1.348 | 364 | 130 | 17 | 25 | － | 49 | 5 | 54 | 63 | 930 | 993 | 286 | 132 | 596 | 496 | 1，092 | 20 | 10 | 6，273 |
| 1983 | 911 | 513 | 1，424 | 1 | 957 | 593 | 1.550 | 507 | 166 | 32 | 28 | － | 51 | 7 | 58 | 150 | 1.506 | 1,656 | 429 | 187 | 672 | 549 | 1,221 | 16 | 23 | 7，298 |
| 1984 | 645 | 467 | 1.112 | 2 | 995 | 628 | 1,623 | 593 | 139 | 20 | 40 | － | 37 | 9 | 46 | 101 | 728 | 829 | 345 | 78 | 304 | 509 | 1，013 | 25 | 18 | 5，883 |
| 1985 | 540 | 593 | 1.133 | 2 | 923 | 638 | 1.561 | 659 | 162 | 53 | 45 | － | 38 | 11 | 49 | 100 | 1.495 | 1.595 | 361 | 98 | 514 | 399 | 913 | 22 | 13 | 6，668 |
| 1986 | 779 | 780 | 1.559 | 2 | 1，042 | 536 | 1.598 | 608 | 232 | 59 | 54 | － | 25 | 12 | 37 | 136 | 1.594 | 1，730 | 430 | 109 | 745 | 526 | 1.271 | 28 | 27 | 7，744 |
| 1987 | 951 | 833 | 1，784 | 1 | 894 | 491 | 1.385 | 364 | 181 | 40 | 47 | － | 34 | 15 | 49 | 127 | 1.112 | 1，239 | 302 | 56 | 503 | 419 | 922 | 27 | 18 | 6，615 |
| 1988 | 633 | 677 | 1,310 | 1 | ${ }_{6} 66$ | 420 | 1，076 | 420 | 217 | 180 | 40 |  | 27 | 9 | 36 | 141 | 1，733 | 1，874 | 395 | 114 | 501 | 381 | 882 | 32 | 18 | 6，593 |
| 1989 | 390 | 349 | 1，139 | 2 | 469 | 436 | 905 | 364 | 141 | 136 | 29 | － | 33 | 19 | 52 | 132 | 947 | 1，079 | 296 | 142 | 464 | 431 | 893 | 14 | 7 | 3，201 |
| 1990 | 486 | 425 | 911 | 2 | 545 | 385 | 930 | 313 | 146 | 280 | 33 | 13 | 41 | 19 | 60 | ． | ． | 567 | 338 | 94 | 423 | 201 | 624 | 15 | 7 | 4.333 |

## Table 2.1.1.2. Continued.

| Year | NAC Area |  |  |  | NEAC (N. Area) |  |  |  |  |  |  |  |  |  |  | NEAC (S. Area) |  |  |  |  |  |  |  |  |  | $\begin{gathered} \text { Total } \\ \mathrm{T} \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Canada (1) |  |  | $\begin{gathered} \text { USA } \\ \text { T } \end{gathered}$ | Norway (2) |  |  | $\begin{gathered} \text { Russia } \\ \left.-\quad \begin{array}{c} \text { (3) } \\ \hline \end{array}\right) \end{gathered}$ | Iceland |  | Sweden |  | Finland |  |  | $\begin{aligned} & \text { Ireland } \\ & (4,5) \end{aligned}$ |  |  | $\begin{gathered} \text { UK } \\ (\text { E\&W) } \\ -T \end{gathered}$ | $\begin{gathered} \mathrm{UK}(\mathrm{~N} . \mathrm{I}) \\ (4,6) \\ \mathrm{T} \end{gathered}$ | UK(Scotland) |  |  | FranceT13 | Spain <br> T |  |
|  |  |  |  | wild |  |  |  | Ranch | (West) | Denmark |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Lg | Sm | T |  | s | G | T |  | T | T | T | T | s | G | T | s | G | T |  |  | s | G | T |  |  |  |
| 1991 | 370 | 341 | 711 |  | 1 | 535 | 342 |  | 876 | 215 | 129 | 346 | 38 | 3 | 53 | 17 | 70 | - | . | 404 | 200 | 55 | 285 |  |  | 177 | 462 | 3,534 |
| 1992 | 323 | 199 | 522 | 1 | 566 | 301 | 867 | 167 | 174 | 462 | 49 | 10 | 49 | 28 | 77 | - | - | 630 | 171 | 91 | 361 | 238 | 599 | 20 | 11 | 3,851 |
| 1993 | 214 | 159 | 373 | 1 | 611 | 312 | 923 | 139 | 157 | 499 | 56 | 9 | 53 | 17 | 70 | - | - | 541 | 248 | 83 | 320 | 227 | 547 | 16 | 8 | 3,670 |
| 1994 | 216 | 139 | 355 | 0 | 581 | 415 | 996 | 141 | 136 | 313 | 44 | 6 | 38 | 11 | 49 | - | - | 804 | 324 | 91 | 400 | 248 | 648 | 18 | 10 | 3,934 |
| 1995 | 153 | 107 | 260 | 0 | 590 | 249 | 839 | 128 | 146 | 303 | 37 | 3 | 37 | 11 | 48 | - | - | 790 | 295 | 83 | 364 | 224 | 588 | 10 | 9 | 3,538 |
| 1996 | 154 | 138 | 292 | 0 | 571 | 215 | 787 | 131 | 118 | 243 | 33 | 2 | 24 | 20 | 44 | - | - | 685 | 183 | 77 | 267 | 160 | 427 | 13 | 7 | 3,042 |
| 1997 | 126 | 103 | 229 | 0 | 389 | 241 | 630 | 111 | 97 | 59 | 19 | 1 | 30 | 15 | 45 | - | - | 570 | 142 | 93 | 182 | 114 | 296 | 8 | 3 | 2,303 |
| 1998 | 70 | 87 | 157 | 0 | 445 | 296 | 740 | 131 | 119 | 46 | 15 | 1 | 29 | 19 | 48 | - | - | 624 | 123 | 78 | 162 | 121 | 283 | 8 | 4 | 2,376 |
| 1999 | 64 | 88 | 152 | 0 | 493 | 318 | 811 | 103 | 111 | 35 | 16 | 1 | 29 | 33 | 62 | - | - | 515 | 150 | 53 | 142 | 57 | 199 | 11 | 6 | 2,225 |
| 2000 | 58 | 95 | 153 | 0 | 673 | 504 | 1,176 | 124 | 73 | 11 | 33 | 5 | 56 | 39 | 95 | - | - | 621 | 219 | 78 | 161 | 114 | 275 | 11 | 7 | 2,882 |
| 2001 | 61 | 86 | 148 | 0 | 850 | 417 | 1,267 | 114 | 74 | 14 | 33 | 6 | 105 | 21 | 126 | - | - | 730 | 184 | 53 | 150 | 101 | 251 | 11 | 13 | 3,024 |
| 2002 | 49 | 99 | 148 | 0 | 770 | 249 | 1,019 | 118 | 90 | 7 | 28 | 5 | 81 | 12 | 93 | - | - | 682 | 161 | 81 | 118 | 73 | 191 | 11 | 9 | 2,643 |
| 2003 | 60 | 81 | 141 | 0 | 708 | 363 | 1,071 | 107 | 99 | 11 | 25 | 4 | 63 | 15 | 78 | - | - | 551 | 89 | 56 | 122 | 71 | 193 | 13 | 7 | 2,444 |
| 2004 | 68 | 94 | 161 | 0 | 577 | 207 | 784 | 82 | 111 | 18 | 19 | 4 | 32 | 7 | 39 | - | - | 489 | 111 | 48 | 159 | 88 | 247 | 19 | 7 | 2,140 |
| 2005 | 56 | 83 | 139 | 0 | 581 | 307 | 888 | 82 | 129 | 21 | 15 | 8 | 31 | 16 | 47 | - | - | 422 | 97 | 52 | 126 | 91 | 217 | 11 | 13 | 2,139 |
| 2006 | 55 | 82 | 137 | 0 | 671 | 261 | 932 | 91 | 93 | 17 | 14 | 2 | 38 | 29 | 67 | - | - | 326 | 80 | 29 | 118 | 75 | 193 | 13 | 11 | 2,003 |
| 2007 | 49 | 63 | 112 | 0 | 627 | 140 | 767 | 63 | 93 | 36 | 16 | 3 | 52 | 6 | 58 | - | - | 85 | 67 | 30 | 100 | 71 | 171 | 11 | 9 | 1,521 |
| 2008 | 57 | 100 | 157 | 0 | 637 | 170 | 807 | 73 | 132 | 69 | 18 | 9 | 65 | 6 | 71 | - | - | 89 | 64 | 21 | 110 | 51 | 161 | 12 | 9 | 1,691 |
| 2009 | 52 | 74 | 126 | 0 | 460 | 135 | 595 | 71 | 122 | 44 | 17 | 8 | 21 | 15 | 36 | - | - | 68 | 54 | 17 | 83 | 37 | 121 | 4 | 2 | 1,284 |
| 2010 | 53 | 100 | 153 | 0 | 458 | 184 | 642 | 88 | 124 | 36 | 22 | 13 | - | - | 49 | - | - | 99 | 109 | 12 | 111 | 69 | 180 | 10 | 2 | 1,538 |
| 2011 | 69 | 110 | 179 | 0 | 556 | 140 | 696 | 89 | 98 | 30 | 39 | 13 | $\cdot$ | $\cdot$ | 44 | - | $\cdot$ | 87 | 136 | 10 | 126 | 33 | 159 | 11 | 7 | 1,598 |
| 2012 | 55 | so | 135 | 0 | 534 | 162 | 696 | 82 | 53 | 12 | 30 | 2 | . | . | 64 | - | . | 88 | 57 | 9 | 88 | 42 | 130 | 10 | 8 | 1,375 |
| Average 2007-2011 2002-2011 | 56 57 | 89 89 | 145 145 | ${ }_{0}^{0}$ | 548 605 | 154 216 | 701 820 | - $\begin{aligned} & 77 \\ & 86\end{aligned}$ | 114 109 | - ${ }^{43}$ | $\cdots{ }^{22}$ | ' 9 | . | . | 52 58 | . | . | 85 290 | $\begin{array}{r} \\ \hline\end{array} 8$ |  | 106 117 | 52 66 | 158 183 | ? $\begin{gathered}9 \\ 11\end{gathered}$ | 6 8 | 1526 1900 |

1. Includes estimates of some local sales, and, prior to 1984, by-catch.
2. Improved reporting of rod catches in 1994 and data derived from carcase tagging and log books from 2002.
3. Before 1966 , sea trout and sea charr included ( $5 \%$ of total).
4. Angling catch (derived from carcase tagging and log books) first included in 2002 .
5. Catch on River Foyle allocated $50 \%$ Ireland and $50 \% \mathrm{~N}$. Ireland.

Table 2.1.2.1. Numbers of fish caught and released in rod fisheries along with the $\%$ of the total rod catch (released + retained) for countries in the North Atlantic where records are available, 1991-2012. Figures for 2012 are provisional.

| Year | Canada ${ }^{4}$ |  | USA |  | Iceland |  | Russia ${ }^{1}$ |  | UK (E\&W) |  | UK (Scotland) |  | Ireland |  | UK (N Ireland) ${ }^{2}$ |  | Denmark |  | Norway ${ }^{3}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total | $\begin{aligned} & \% \text { of total } \\ & \text { rod } \\ & \text { catch } \end{aligned}$ | Total | $\begin{gathered} \% \text { of total } \\ \text { rod } \\ \text { catch } \end{gathered}$ | Total | $\begin{aligned} & \hline \% \text { of total } \\ & \text { rod } \\ & \text { catch } \end{aligned}$ | Total | $\begin{gathered} \% \text { of total } \\ \text { rod } \\ \text { catch } \end{gathered}$ | Total | $\begin{aligned} & \% \text { of total } \\ & \text { rod } \\ & \text { catch } \end{aligned}$ | Total | $\%$ of total rod catch | Total | $\begin{aligned} & \% \text { of total } \\ & \text { rod } \\ & \text { catch } \end{aligned}$ | Total | $\begin{aligned} & \% \text { of total } \\ & \text { rod } \\ & \text { catch } \end{aligned}$ | Total | $\begin{gathered} \% \text { of total } \\ \text { rod } \\ \text { catch } \end{gathered}$ | Total | $\begin{gathered} \% \text { of total } \\ \text { rod } \\ \text { catch } \end{gathered}$ |
| 1991 | 22,167 | 28 | 239 | 50 |  |  | 3,211 | 51 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | 37,803 | 29 | 407 | 67 |  |  | 10,120 | 73 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1993 | 44,803 | 36 | 507 | 77 |  |  | 11,246 | 82 | 1,448 | 10 |  |  |  |  |  |  |  |  |  |  |
| 1994 | 52,887 | 43 | 249 | 95 |  |  | 12,056 | 83 | 3,227 | 13 | 6,595 | 8 |  |  |  |  |  |  |  |  |
| 1995 | 46,029 | 46 | 370 | 100 |  |  | 11,904 | 84 | 3,189 | 20 | 12,151 | 14 |  |  |  |  |  |  |  |  |
| 1996 | 52,166 | 41 | 542 | 100 | 669 | 2 | 10,745 | 73 | 3,428 | 20 | 10,413 | 15 |  |  |  |  |  |  |  |  |
| 1997 | 50,009 | 50 | 333 | 100 | 1,558 | 5 | 14,823 | 87 | 3,132 | 24 | 10,965 | 18 |  |  |  |  |  |  |  |  |
| 1998 | 56,289 | 53 | 273 | 100 | 2,826 | 7 | 12,776 | 81 | 4,378 | 30 | 13,464 | 18 |  |  |  |  |  |  |  |  |
| 1999 | 48,720 | 50 | 211 | 100 | 3,055 | 10 | 11,450 | 77 | 4,382 | 42 | 14,846 | 28 |  |  |  |  |  |  |  |  |
| 2000 | 64,482 | 56 | 0 | - | 2,918 | 11 | 12,914 | 74 | 5,959 | 40 | 21,072 | 32 |  |  |  |  |  |  |  |  |
| 2001 | 59,387 | 55 | 0 | - | 3,611 | 12 | 16,945 | 76 | 4,869 | 41 | 27,724 | 38 |  |  |  |  |  |  |  |  |
| 2002 | 50,924 | 52 | 0 | - | 5,985 | 18 | 25,248 | 80 | 5,910 | 47 | 24,058 | 42 |  |  |  |  |  |  |  |  |
| 2003 | 53,645 | 55 | 0 | - | 5,361 | 16 | 33,862 | 81 | 4,943 | 53 | 29,170 | 55 |  |  |  |  |  |  |  |  |
| 2004 | 62,316 | 57 | 0 | - | 7,362 | 16 | 24,679 | 76 | 11,516 | 46 | 46,279 | 50 |  |  |  |  | 255 | 19 |  |  |
| 2005 | 63,005 | 62 | 0 | - | 9,224 | 17 | 23,592 | 87 | 10,554 | 54 | 46,165 | 55 | 2,553 | 12 |  |  | 606 | 27 |  |  |
| 2006 | 60,486 | 62 | 1 | 100 | 8,735 | 19 | 33,380 | 82 | 9,955 | 55 | 47,669 | 55 | 5,409 | 22 | 302 | 18 | 794 | 65 |  |  |
| 2007 | 41,192 | 58 | 3 | 100 | 9,691 | 18 | 44,341 | 90 | 9,942 | 53 | 55,660 | 61 | 13,125 | 40 | 470 | 16 | 959 | 57 |  |  |
| 2008 | 54,887 | 53 | 61 | 100 | 17,178 | 20 | 41,881 | 86 | 11,918 | 54 | 53,347 | 62 | 13,312 | 37 | 648 | 20 | 2,033 | 71 | 5,512 | 5 |
| 2009 | 52,151 | 59 | 0 | - | 17,514 | 24 | - | - | 8,397 | 57 | 48,418 | 67 | 10,265 | 37 | 847 | 21 | 1,709 | 53 | 6,696 | 6 |
| 2010 | 55,895 | 53 | 0 | - | 21,476 | 29 | 14,585 | 56 | 13,958 | 59 | 78,304 | 70 | 15,136 | 40 | 823 | 25 | 2,512 | 60 | 15,041 | 12 |
| 2011 | 71,358 | 57 | 0 | - | 18,593 | 32 | - | - | 13,471 | 61 | 64,669 | 73 | 12,753 | 39 | 1,197 | 36 | 2,153 | 55 | 14,303 | 12 |
| 2012 | 50,811 | 57 | 0 | - | 7,963 | 28 | 4,743 | 43 | 10,967 | 64 | 66,250 | 74 | 11,891 | 35 | 5,014 | 59 | 2,153 | 55 | 18,611 | 14 |
| $\begin{array}{\|l\|} \hline \text { 5-yr mean } \\ \text { 2007-2011 } \\ \hline \end{array}$ | 55,096 | 56 | 0 |  | 16,890 | 24 |  |  | ' 11,537 | $\cdots 57$ | '60,080 | ' 66 | ' 12,918 | ' 39 | $\checkmark 797$ | 24 | ' 1,873 | 59 | ' 10,388 | 9 |
| \% change on 5-year mean | -8 | +2 |  |  | -53 | +14 |  |  | -5 | +13 | +10 | +12 | -8 | -9 | +529 | +151 | +15 | -7 | +79 | +63 |

Key: $\quad{ }^{1}$ No data were provided by the authorities for 2009 nor for 2011 and data for 2010 and 2012 were incomplete, however catch-and-release is understood to have remained at similar high levels.
${ }^{2}$ Data for 2006-2009 is for the DCAL area only; the figures from 2010 are a total for UK (N.Ireland).
${ }^{3}$ The statistics were collected on a voluntary basis, the numbers reported must be viewed as a minimum
${ }^{4}$ Released fish in the kelt fishery of New Brunswick are not included in the totals for Canada.

Table 2.1.3.1. Estimates of unreported catches (tonnes round fresh weight) by various methods within national EEZs in the Northeast Atlantic, North American and West Greenland Commissions of NASCO, 1987-2012.

| Year | North-East <br> Atlantic | North-America | West <br> Greenland | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1987 | 2,554 | 234 | - | 2,788 |
| 1988 | 3,087 | 161 | - | 3,248 |
| 1989 | 2,103 | 174 | - | 2,277 |
| 1990 | 1,779 | 111 | - | 1,890 |
| 1991 | 1,555 | 127 | - | 1,682 |
| 1992 | 1,825 | 137 | - | 1,962 |
| 1993 | 1,471 | 161 | $<12$ | 1,644 |
| 1994 | 1,157 | 107 | $<12$ | 1,276 |
| 1995 | 942 | 98 | 20 | 1,060 |
| 1996 | 947 | 156 | 20 | 1,123 |
| 1997 | 732 | 90 | 5 | 827 |
| 1998 | 1,108 | 91 | 11 | 1,210 |
| 1999 | 887 | 133 | 12.5 | 1,032 |
| 2000 | 1,135 | 124 | 10 | 1,269 |
| 2001 | 1,089 | 81 | 10 | 1,180 |
| 2002 | 946 | 83 | 10 | 1,039 |
| 2003 | 719 | 118 | 10 | 847 |
| 2004 | 575 | 101 | 10 | 686 |
| 2005 | 605 | 85 | 10 | 700 |
| 2006 | 604 | 56 | 10 | 670 |
| 2007 | 465 | - | 10 | 475 |
| 2008 | 433 | - | 10 | 443 |
| 2009 | 317 | 16 | 10 | 343 |
| 2010 | 357 | 26 | 10 | 393 |
| 2011 | 382 | 29 | 10 | 421 |
| 2012 | 363 | 31 | 10 | 403 |
| $\begin{gathered} \text { Mean } \\ 2007-2011 \\ \hline \end{gathered}$ | 391 |  | 10 | 415 |

Notes:
There were no estimates available for Canada in 2007-08 and estimates for 2009-10 are incomplete. No estimates have been available for Russia since 2008.
Unreported catch estimates are not provided for Spain and St. Pierre et Miquelon.

Table 2.1.3.2. Estimates of unreported catches by various methods in tonnes by country within national EEZs in the Northeast Atlantic, North American and West Greenland Commissions of NASCO, 2012.

| Commission Area | Country | Unreported Catch t | Unreported as \% of Total North Atlantic Catch (Unreported + Reported) | Unreported as \% of Total National Catch (Unreported + Reported) |
| :---: | :---: | :---: | :---: | :---: |
| NEAC | Denmark | 6 | 0.3 | 77 |
| NEAC | Finland | 7 | 0.4 | 10 |
| NEAC | Iceland | 5 | 0.3 | 8 |
| NEAC | Ireland | 9 | 0.5 | 9 |
| NEAC | Norway | 298 | 16.4 | 30 |
| NEAC | Sweden | 3 | 0.2 | 9 |
| NEAC | France | 2 | 0.1 | 14 |
| NEAC | UK (E \& W) | 15 | 0.8 | 21 |
| NEAC | UK (N.Ireland) | 0 | 0.0 | 2 |
| NEAC | UK (Scotland) | 18 | 1.0 | 12 |
| NAC | USA | 0 | 0.0 | 0 |
| NAC | Canada | 31 | 1.7 | 18 |
| WGC | West Greenland | 10 | 0.6 | 23 |
|  | Total Unreported Catch * | 403 | 22.3 |  |
|  | Total Reported Catch of North Atlantic salmon | 1,409 |  |  |

* No unreported catch estimate available for Russia in 2012.

Unreported catch estimates not provided for Spain \& St. Pierre et Miquelon

Table 2.2.1.1. Production of farmed salmon in the North Atlantic area and in areas other than the North Atlantic (in tonnes round fresh weight), 1980-2012.

| Year | North Atlantic Area |  |  |  |  |  |  |  |  |  | Outside the North Atlantic Area |  |  |  |  |  |  | World-wide |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Norway | $\begin{gathered} \mathrm{UK} \\ \text { (Scot.) } \end{gathered}$ | Faroes | Canada | Ireland | USA | Iceland | $\begin{gathered} \hline \text { UK } \\ \text { (N.Ire.) } \end{gathered}$ | Russia | Total | Chile | West <br> Coast <br> USA | West Coast Canada | Australia | Turkey | Other | Total | Total |
| 1980 | 4,153 | 598 | 0 | 11 | 21 | 0 | 0 | 0 | 0 | 4,783 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4,783 |
| 1981 | 8,422 | 1,133 | 0 | 21 | 35 | 0 | 0 | 0 | 0 | 9,611 | 0 | 0 | 0 | 0 | 0 | 0 |  | 9,611 |
| 1982 | 10,266 | 2,152 | 70 | 38 | 100 | 0 | 0 | 0 | 0 | 12,626 | 0 | 0 | 0 | 0 | 0 | 0 |  | 12,626 |
| 1983 | 17,000 | 2,536 | 110 | 69 | 257 | 0 | 0 | 0 | 0 | 19,972 | 0 | 0 | 0 | 0 | 0 | 0 |  | 19,972 |
| 1984 | 22,300 | 3,912 | 120 | 227 | 385 | 0 | 0 | 0 | 0 | 26,944 | 0 | 0 | 0 | 0 | 0 | 0 |  | 26,944 |
| 1985 | 28,655 | 6,921 | 470 | 359 | 700 | 0 | 91 | 0 | 0 | 37,196 | 0 | 0 | 0 | 0 | 0 | 0 |  | 37,196 |
| 1986 | 45,675 | 10,337 | 1,370 | 672 | 1,215 | 0 | 123 | 0 | 0 | 59,392 | 0 | 0 | 0 | 20 | 0 | 0 |  | 59,392 |
| 1987 | 47,417 | 12,721 | 3,530 | 1,334 | 2,232 | 365 | 490 | 0 | 0 | 68,089 | 3 | 0 | 0 | 50 | 0 | 0 |  | 68,142 |
| 1988 | 80,371 | 17,951 | 3,300 | 3,542 | 4,700 | 455 | 1,053 | 0 | 0 | 111,372 | 174 | 0 | 0 | 250 | , | 0 | 424 | 111,796 |
| 1989 | 124,000 | 28,553 | 8,000 | 5,865 | 5,063 | 905 | 1,480 | 0 | 0 | 173,866 | 1,864 | 1,100 | 1,000 | 400 | 0 | 700 | 5,064 | 178,930 |
| 1990 | 165,000 | 32,351 | 13,000 | 7,810 | 5,983 | 2,086 | 2,800 | $<100$ | 5 | 229,035 | 9,500 | 700 | 1,700 | 1,700 | 0 | 800 | 14,400 | 243,435 |
| 1991 | 155,000 | 40,593 | 15,000 | 9,395 | 9,483 | 4,560 | 2,680 | 100 | 0 | 236,811 | 14,991 | 2,000 | 3,500 | 2,700 | 0 | 1,400 | 24,591 | 261,402 |
| 1992 | 140,000 | 36,101 | 17,000 | 10,380 | 9,231 | 5,850 | 2,100 | 200 | 0 | 220,862 | 23,769 | 4,900 | 6,600 | 2,500 | 0 | 400 | 38,169 | 259,031 |
| 1993 | 170,000 | 48,691 | 16,000 | 11,115 | 12,366 | 6,755 | 2,348 | $<100$ | 0 | 267,275 | 29,248 | 4,200 | 12,000 | 4,500 | 1,000 | 400 | 51,348 | 318,623 |
| 1994 | 204,686 | 64,066 | 14,789 | 12,441 | 11,616 | 6,130 | 2,588 | $<100$ | 0 | 316,316 | 34,077 | 5,000 | 16,100 | 5,000 | 1,000 | 800 | 61,977 | 378,293 |
| 1995 | 261,522 | 70,060 | 9,000 | 12,550 | 11,811 | 10,020 | 2,880 | 259 | 0 | 378,102 | 41,093 | 5,000 | 16,000 | 6,000 | 1,000 | 0 | 69,093 | 447,195 |
| 1996 | 297,557 | 83,121 | 18,600 | 17,715 | 14,025 | 10,010 | 2,772 | 338 | 0 | 444,138 | 69,960 | 5,200 | 17,000 | 7,500 | 1,000 | 600 | 101,260 | 545,398 |
| 1997 | 332,581 | 99,197 | 22,205 | 19,354 | 14,025 | 13,222 | 2,554 | 225 | 0 | 503,363 | 87,700 | 6,000 | 28,751 | 9,000 | 1,000 | 900 | 133,351 | 636,714 |
| 1998 | 361,879 | 110,784 | 20,362 | 16,418 | 14,860 | 13,222 | 2,686 | 114 | 0 | 540,325 | 125,000 | 3,000 | 33,100 | 7,068 | 1,000 | 400 | 169,568 | 709,893 |
| 1999 | 425,154 | 126,686 | 37,000 | 23,370 | 18,000 | 12,246 | 2,900 | 234 | 0 | 645,590 | 150,000 | 5,000 | 38,800 | 9,195 | - | 500 | 203,495 | 849,085 |
| 2000 | 440,861 | 128,959 | 32,000 | 33,195 | 17,648 | 16,461 | 2,600 | 250 | 0 | 671,974 | 176,000 | 5,670 | 49,000 | 12,003 |  | 500 | 243,173 | 915,147 |
| 2001 | 436,103 | 138,519 | 46,014 | 36,514 | 23,312 | 13,202 | 2,645 |  | 0 | 696,309 | 200,000 | 5,443 | 68,000 | 13,815 | 0 | 500 | 287,758 | 984,067 |
| 2002 | 462,495 | 145,609 | 45,150 | 40,851 | 22,294 | 6,798 | 1,471 |  | 0 | 724,668 | 273,000 | 5,948 | 84,200 | 14,699 | 0 | 1,000 | 378,847 | 1,103,515 |
| 2003 | 509,544 | 176,596 | 52,526 | 38,680 | 16,347 | 6,007 | 3,710 |  | 298 | 803,708 | 261,000 | 10,329 | 65,411 | 13,324 | 0 | 1,000 | 351,064 | 1,154,772 |
| 2004 | 563,914 | 158,099 | 40,492 | 37,280 | 14,067 | 8,515 | 6,620 | - | 203 | 829,190 | 261,000 | 6,659 | 55,646 | 14,317 | 0 | 1,000 | 338,622 | 1,167,812 |
| 2005 | 586,512 | 129,588 | 18,962 | 45,891 | 13,764 | 5,263 | 6,300 | . | 179 | 806,459 | 385,000 | 6,123 | 63,369 | 16,827 | 0 | 1,000 | 472,319 | 1,278,778 |
| 2006 | 629,888 | 131,847 | 11,905 | 47,880 | 11,000 | 4,674 | 5,745 | - | 229 | 843,168 | 370,000 | 5,823 | 70,181 | 22,417 | 0 | 1,000 | 469,421 | 1,312,589 |
| 2007 | 744,222 | 129,930 | 22,305 | 36,368 | 9,923 | 2,715 | 1,158 | - | 280 | 946,901 | 371,809 | 6,261 | 70,998 | 23,982 | 0 | 1,000 | 474,050 | 1,420,951 |
| 2008 | 737,694 | 128,606 | 36,000 | 39,687 | 11,000 | 9,014 | 330 | - | 380 | 962,711 | 393,000 | 6,261 | 73,265 | 26,173 | 0 | 1,000 | 499,699 | 1,462,410 |
| 2009 | 862,908 | 144,247 | 51,500 | 43,101 | 13,000 | 6,028 | 742 | - | 55 | 1,121,581 | 200,000 | 7,930 | 68,662 | 32,819 | 0 | 1,000 | 310,411 | 1,431,992 |
| 2010 | 939,575 | 154,164 | 45,396 | 43,612 | 13,000 | 11,127 | 1,068 | - | 1,400 | 1,209,342 | 81,000 | 7,930 | 70,831 | 30,264 | 0 | 1,000 | 191,025 | 1,400,367 |
| 2011 | 1,065,974 | 158,018 | 60,500 | 41,448 | 14,000 | - | 1,083 | - | 4,000 | 1,345,023 | 385,000 | 8,014 | 74,880 | 35,685 | 0 | 1,000 | 504,579 | 1,849,602 |
| 2012 | 1,147,745 | 158,026 | 76,595 | 41,448 | 15,000 | . | 2,923 | . | 8,000 | 1,449,737 | 385,000 | 7,131 | 74,880 | 43,249 | 0 | 1,000 | 511,260 | 1,960,997 |
| $\begin{array}{\|l\|} \hline \text { 5-yr mean } \\ \text { 2007-2011 } \\ \hline \end{array}$ | 870,075 | 142,993 | 43,140 | 40,843 | 12,185 |  | 876 |  | 1,223 | 1,117,112 | 286,162 | 7,279 | 71,727 | 29,785 | 0 | 1,000 | 395,953 | 1,513,064 |
| \% change on 5 -year mean | +32 | +11 | +78 | +1 | +23 |  | +234 |  | +554 | +30 | +35 | -2 | +4 | +45 |  | 0 | +29 | +30 |

Notes: Data for 2012 are provisional for many countries.
here production figures were not avila
os Coal a
Australia = Tasmania. This is mostly Atlantic salmon, but includes a small component of trout
Australia $=T$ asmarnia. This is mostly Athantic salmon, but includes a small componient of trout
Source of production figures for non-Atlantic areas: miscellaneous fishing publications \& Goverment reports
Other' includes South Korea \& China
Data for UK (N. Ireland) since 2001 and data for USA since 2011 are not publicly available.

Table 2.2.2.1. Production of ranched salmon in the North Atlantic (tonnes round fresh weight), 19802012.

| Year | Iceland (1) | Ireland (2) | UK(N.Ireland) <br> River Bush (2,3) | Norway various facilities (2) | Total production |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 8.0 |  |  |  | 8 |
| 1981 | 16.0 |  |  |  | 16 |
| 1982 | 17.0 |  |  |  | 17 |
| 1983 | 32.0 |  |  |  | 32 |
| 1984 | 20.0 |  |  |  | 20 |
| 1985 | 55.0 | 16.0 | 17.0 |  | 88 |
| 1986 | 59.0 | 14.3 | 22.0 |  | 95 |
| 1987 | 40.0 | 4.6 | 7.0 |  | 52 |
| 1988 | 180.0 | 7.1 | 12.0 | 4.0 | 203 |
| 1989 | 136.0 | 12.4 | 17.0 | 3.0 | 168 |
| 1990 | 285.1 | 7.8 | 5.0 | 6.2 | 304 |
| 1991 | 346.1 | 2.3 | 4.0 | 5.5 | 358 |
| 1992 | 462.1 | 13.1 | 11.0 | 10.3 | 497 |
| 1993 | 499.3 | 9.9 | 8.0 | 7.0 | 524 |
| 1994 | 312.8 | 13.2 | 0.4 | 10.0 | 336 |
| 1995 | 302.7 | 19.0 | 1.2 | 2.0 | 325 |
| 1996 | 243.0 | 9.2 | 3.0 | 8.0 | 263 |
| 1997 | 59.4 | 6.1 | 2.8 | 2.0 | 70 |
| 1998 | 45.5 | 11.0 | 1.0 | 1.0 | 59 |
| 1999 | 35.3 | 4.3 | 1.4 | 1.0 | 42 |
| 2000 | 11.3 | 9.3 | 3.5 | 1.0 | 25 |
| 2001 | 13.9 | 10.7 | 2.8 | 1.0 | 28 |
| 2002 | 6.7 | 6.9 | 2.4 | 1.0 | 17 |
| 2003 | 11.1 | 5.4 | 0.6 | 1.0 | 18 |
| 2004 | 18.1 | 10.4 | 0.4 | 1.0 | 30 |
| 2005 | 20.5 | 5.3 | 1.7 | 1.0 | 29 |
| 2006 | 17.2 | 5.8 | 1.3 | 1.0 | 25 |
| 2007 | 35.5 | 3.1 | 0.3 | 0.5 | 39 |
| 2008 | 68.6 | 4.4 | - | 0.5 | 74 |
| 2009 | 44.3 | 1.1 | - | - | 45 |
| 2010 | 42.3 | 2.5 | - | - | 45 |
| 2011 | 30.2 | 3.2 | - | - | 33 |
| 2012 | 11.7 | - | - | - | 12 |
| 5-yr mean 2007-2011 | 44.2 | 2.9 |  |  | 47 |
| \% change on 5 year mean | -74 |  |  |  | -75 |

1 From 1990, catch includes fish ranched for both commercial and angling purposes.
2 Total yield in homewater fisheries and rivers, estimate for 2012 is not available.
3 The proportion of ranched fish was not assessed between 2008 and 2011 due to a lack of microtag returns.

Table 2.3.8.2.1. Summary of prevalence ( $\%$ of sampled fish with sea lice) of sea lice on Atlantic salmon sampled at West Greenland in 2009 to 2011 and from the Miramichi River (Canada) in 2005 to 2011.

| Lice CATEGORIES | West Grenland |  |  | MIRAMICHI RIVER |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2009 | 2010 | 2011 | mean | $2005-2011$ |
|  | $24.9 \%$ | $27.1 \%$ | $36.7 \%$ | $29.6 \%$ | $59.1 \%$ |
| $1-5$ | $57.3 \%$ | $55.7 \%$ | $50.5 \%$ | $54.5 \%$ | $21.3 \%$ |
| $6-10$ | $12.4 \%$ | $13.1 \%$ | $9.3 \%$ | $11.6 \%$ | $14.6 \%$ |
| $11-15$ | $4.4 \%$ | $2.3 \%$ | $2.6 \%$ | $3.1 \%$ |  |
| $16-20$ | $0.8 \%$ | $1.1 \%$ | $0.7 \%$ | $0.9 \%$ | $4.8 \%$ |
| $21-50$ | $0.3 \%$ | $0.6 \%$ | $0.2 \%$ | $0.4 \%$ |  |
| $>50$ | 0 | 0 | 0 | 0 | $0.3 \%$ |

Table 2.5.2.1a. Distribution of exotic salmonids in NEAC Northern area.

| Northern NEAd | Rainbow trout | American Brook Trout | Lake Trout | Pink salmon | Coho Salmon | Chinook salmon | Landlocked Atlantic salmon | Brown trout |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Country | Oncorhynchus mykiss | Salvelinus fontinalis | Salvelinus namaycush | Oncorhynchus gorbuscha | Oncorhynchus kisutch | Oncorhynchus <br> tshawytscha | salmo salar | Salmo trutta |
| Russia | Widely used in aquaculture in Karelia Republic, including farming in the White Sea. One freshwater farm in Tuloma River (Barents Sea basin) and a number of farms in inland waters of the Murmansk region. Limited production of rainbow trout in the lakes of the Archangelsk region. No population established in wild. Limited evidence of escapees. | No history of use | No history of use | Pink salmon was successfully introduced to salmon rivers of Murmansk and Archangelsk regions from the Russia's Far East from 1950 to 1990 . Since 2003 no juvenile releases from hatcheries occured but pink salmon spawned naturally in salmon rivers. <br> The commercial fishery for pink salmon is conducted in the coastal areas of the White Sea with the same gears and in the same period as for Atlantic salmon. The total declared pink salmon catch in 2009 was 139 t , twice as much as a declared Atlantic salmon catch. | Experiments with Coho salmon introductions were conducted in Murmansk region in 1930's. No evidence of adult returns. | Not present | There are wild landlocked populations of Atlantic salmon in big lakes of Karelia. Salmon population of Shuya River of Onega lake is widely used in recreational fisheries. | Native species. Widely used in commercial and recreational fisheries. |
| Iceland | Introduced in the mid 1950's. Used in aquaculture in land based facitlities and in sea-cages. Production increasing to $400 \mathrm{t} / \mathrm{ye}$ ear. No evidence of impact. | Not present | Not present | Caught annually in Icelandic rivers, since the 1960 's, in low numbers $5-30$ /year. No evidence of impact. | Not present | Not present | Not present | Widely distrubuted native species, both resident and sea-run. |
| Finland | Not present | Not present | Not present | Caught occasionally in rivers. Not considered a threat currently | Not present |  | Not present | Native and widespread in Finland |
| Norway | In use for $>100$ years. However, only 6 instances of natural production confirmed by mid 1990s, 2 in anadromous parts of rivers (Jonsson et al. 1993a; Hindar et al. 1996; Sægrov, Hindar \& Urdal 1996; Hesthagen \& Sandlund 2007) | Reported from at least 12 lakes in Norway, first reportsin 1974. Illegaly stocked; no natural spreading. Hesthagen \& Sandlund, (2007). | It is at present assumed to be less than 50 populations in Norway, and the number of populations are decreasing. No information on potential | Pink salmon introduced to Russia since 1930s have resulted in catches in Norwegian waters (up to 20t in some years). The species has also now established in 11 rivers in N. Norway (Finnmark) Hesthagen \& Sandlund (2007) | Not present | Not present | Two-three stocks native to Norway | Native and widespread in Norway |
| Sweden | In Sweden since the 1880 s. Occasional natural reproduction is rare (40 known occasions, none of which are on the Swedish west coast). No self-sustaining populations known. | Stocked since the late 1800 s. Established self-reproducing populations (ca 1000) in headwaters (Öhlund et al. 2009), not in river sections with Atlantic salmon. | Stocked in the 1960s- <br> 1970s. Today self- <br> reproducing in 16 <br> lakes (all draining to <br> the Baltic Sea). In <br> some lakes, a fishery <br> has started to <br> deplete the | Not occurring. | Not occurring. | Not occurring. | Two indigenous stocks exist in Lake Vänern (Piccolo et al. 2012). River Klarälven stock with ca 1000 spawners annually, R. <br> Gullspảngsälven with less than 100 spawners. <br> Stocking of reared salmon | Indigenous and widespread. |

N.B. Table excludes grayling (Thymallus thymallus). This species is native to some NEAC areas, but not UK (Scotland) where it has been introduced over the past 160 years and established self-maintaining populations in several river systems.

## Table 2.5.2.1a. Distribution of exotic salmonids in NEAC Southern area.

| Southern NEAC | Rainbow trout | American Brook Trout | Lake Trout | Pink salmon | Coho Salmon | Chinook salmon | Landlocked Atlantic salmon | Brown trout |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Country | Oncorhynchus mykiss | Salvelinus fontinalis | Salvelinus namaycush | Oncorhynchus gorbuscha | Oncorhynchus kisutch | $\begin{aligned} & \text { Oncorhynchus } \\ & \text { tshawytscha } \end{aligned}$ | Salmo salar | Salno trutta |
| Ireland | Commonly used for small put and take fisheries. Possibly one spawning population. Not considered a threat | Not present | Not present | Not present | Not present | Not present | Not present | Native species |
| UK (Scotland) ${ }^{1}$ | Reared in many fish farms and stocked into many sport fisheries for $>100$ years. Both stocking and the operation of fish farms is now regulated by Marine Scotland. No naturally self maintaining populations have established. Not currently stocked into rivers, but sometimes occur in rivers as a result of escapes from stocked fisheries or fish farms. | Reared in some fish farms and stocked into some sport fisheries for >100 years. Both stocking and the operation of fish farms is regulated by Marine Scotland. A few naturally self maintaining populations have established. Not currently stocked into rivers, and reports from rivers rare. | No history of presence, and not reared | Some reports of captures in the 1960 s and 1970s (in net fisheries) and since 2003 (mainly by rod). No populations have established and not reared. | No history of presence, and not reared | Not present | Not currently present. | Native to all parts of Scotland. Also reared in many fish farms and stocked into many sport fisheries |
| UK (England \& Wales) | Widely used in aquaculture and sport fisheries for $>150$ years. Only one small population established in wild. Limited evidence of impact. | Limited past use in aquaculture and sport fisheries, but not currently present. No evidence of impact. | No history of use | No history of use | No history of use | No history of use | Limited past use in aquaculture and sport fisheries, but not currently present. No evidence of impact. | Native species. Restrictions apply to release to minimise risk of impacts on existing wild populations |
| UK (N. Ireland) | Used in aquaculture and recreational put-and-take fisheries in lakes. No established self sustaining populations known. Escapees, both resident and sea-run, are occasionally encountered. Limited threat to native salmonid | No history of use | No history of use | No history of use. A few events of this species appearing in the River Bush adult trap have been documented in the recent past. These were probably strays from stocking activities of this species elsewhere in Europe | No history of use | No history of use | No history of use | Native species. |
| France | Widely used in farming and sport fisheries for $>150$ years, found in AS and Sea trout rivers. No population known to be established in wild. Some returns known in river Bresle of large steelhead, coming from rearing of Oregon steelhead strain for aqauculture. G. Euzenat, com pers) | introduced in 1876 and after.In 1930-50 in rivers and lakes in east, central, south-west. Currently present in theses areas. No incidence of reproduction have been investigated in rivers.. | not present | not present | Not present currently. Escapees from hatcheries for aquaculture in 1970s and 80 s in NW, W and SW rivers. Escaping fish investigated in R. Bresle (upper-Normandy) in $83 s$ but very low returns noted. Establishment is not considered possible. Limited by failure to return due to adverse oceanic conditions.(1). No viable hvbridization | Not present | not present | Native species. Wild pop. and stockings. No specific restrictions in salmon rivers. Possible problems with smolt catches by trout anglers in spring. Possible competition for food and space with sea trout in mixed pop.,especially in chalk streams where spawning grounds are limited. |

## Table 2.5.2.1b. Distribution of exotic salmonids in NAC area.

| NAC Area | Rainbow trout | American Brook Trout | Lake Trout | Pink salmon | Coho Salmon | Chinook salmon | Landlocked Atlantic salmon | Brown trout |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Country | Oncorhynchus mykiss | Salvelinus fontinalis | Salvelinus namaycush | Oncorhynchus gorbuscha | Oncorhynchus kisutch | Oncorhynchus tshawytscha | Salmo salar | Salmo trutta |
| Canada - <br> Labrador | Not present. | Native | Native | No history of use, but incidental captures in north Labrador during the late 1970s. | No history of use. | No history of use. | Native | No history of use. |
| Canada - Nfld | First introduced in 1887 with stocking continuing through the first half of the 1900s. Self sustaining populations in eastern Nfld. Numerous occurrences have now been documented from various south and west coast rivers. | Native | Not present. | Introduced to Newfoundland from 19581966. No longer present. | No history of use. | No history of use. | Native | Introductions occurred from 1882-1906. Current distribution is estimated at 68 watersheds in eastern Nfld (2011). |
| Canada Quebec | Introduced from Northern Pacific Ocean starting in 1893 for recreational fishing. Increasing prevalence in Eastern Québec since 1980s. Currently present in about 50 river systems. | Native | Native | Historically introduced in Newfoundland and some migrants were observed in Quebec. No longer present. | Historically introduced in the Great Lakes from 1873. Sometimes observed in the Saint Lawrence River. | Historically introduced in the Great Lakes from 1873. Sometimes observed in the Saint Lawrence River. | Native | Introduced in 1890. Now reproducing naturally in multiple watersheds. |
| Canada - Gulf | First introduced to Prince Edward Island in 1924 and self-sustaining populations now present in 18 watersheds. Observations in numerous other rivers in the region in many years. One reproducing population suspected but not confirmed in a tributary of the Restigouche River, New Brunswick. | Native, widely distributed in the region. | Native, found in a few inland lakes in New Brunswick. | No history of use. | No history of use. | No history of use. | Native in a few inland lakes of New Brunswick | Successfully introduced and self-sustaining populations found in most rivers in Nova Scotia flowing into the Gulf of St. Lawrence. Not reported from New Brunswick or Prince Edward Island rivers. |
| Canada - Scotia Fundy | Introduced from 1899 to early 1970 s for recreational fishing. Self-sustaining populations in two sub-watersheds within the Inner Bay of Fundy New Brunswick. Direct stocking still occurs by the province of Nova Scotia in limited number of watersheds in Nova Scotia. No successful populations established in Nova Scotia. <br> Juvenilesfound in many watersheds in the New Brunswick Bay of Fundy area. | Native | Native | Historically introduced in Newfoundland and some migrants were observed in Nova Scotia. No longer present. | Introduced in northeastern USA in early 1970 s to 1990 s and some successful reproductions occurred in mid-80s in Scotia Fundy. No self sustaining population present. | Historically introduced to New <br> Brunswick. No established populations. | Native. Stocking programs occur in New Brunswick Bay of Fundy rivers. | Introduced in 1920s. Now reproducing naturally in multiple watersheds. Direct stocking still occurs by the province of Nova Scotia in limited number of watersheds in Nova Scotia. Successful populations established in Nova Scotia are rare. |
| USA | Stocked in lakes and rivers, produced naturalized populations, typically introduced in riverine habitat where ATS were extirpated, compete for food and space | Native in US Atlantic salmon habitat | Native to US <br> Atlantic salmon <br> habitat, populations <br> in lakes, limited <br> riverine presence | Stocked in the past in habitat where salmon were extirpated, never produced naturalized populations | Stocked in the past in habitat where salmon were extirpated, never produced naturalized populations |  | widely distributed outside landlocked salmon native range in Northeast US, hybridize with anadromous ATS, use same habitat as spawners and juveniles | Stocked in lakes and rivers, produced naturalized populations, typically introduced in riverine habitat where salmon were extirpated, compete for food and space, hybridize |

Table 2.7.1.1. Summary of Atlantic salmon tagged and marked in 2012 ; 'Hatchery' and 'Wild' refer to smolts and parr; 'Adults' relates to both wild and hatchery-origin fish.

| Country | Orizin | Primary Tag or Mark |  |  | Other Intemal | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Microtag | Extemal mark | Adipose clip |  |  |
| Canada | Hatchery Adult | 0 | 2,164 | 0 | 1,949 | 4,113 |
|  | Hatchery Juverile | 0 | 1,292 | 383,332 | 39 | 384,663 |
|  | Wild Adult ${ }^{\text {2 }}$ | 0 | 2,297 | 0 | 25 | 2,322 |
|  | Wild Juverile ${ }^{2}$ | 0 | 15,098 | 10,661 | 640 | 26,399 |
|  | Total | 0 | 20,851 | 393,993 | 2,653 | 417,497 |
| Dermark | Hatchery Adult | 0 | 0 | 0 | 0 | 0 |
|  | Hatchery Juverile | 118,500 | 0 | 152,600 | 0 | 271,100 |
|  | Wild Adult | 0 | 0 | 0 | 0 | 0 |
|  | Wild Juverile | 0 | 0 | 0 | 0 | 0 |
|  | Total | 118,500 | 0 | 152,600 | 0 | 271,100 |
| France | Hatchery Adult | 0 | 0 | 0 | 0 | 0 |
|  | Hatchery Juverile | 0 | 0 | 619,901 | 0 | 619,901 |
|  | Wild Adalt ${ }^{\text {a }}$ | 31 | 0 | 0 | 279 | 310 |
|  | Wild Juverile | 596 |  | 0 | 2,391 | 2,987 |
|  | Total | 627 | 0 | 619,901 | 2,670 | 623,198 |
| Germany | Hatchery Adult | 0 | 0 | 0 | 0 | 0 |
|  | Hatchery Juverile | 12,000 | 0 | 120,000 | 0 | 132,000 |
|  | Wild Adult | 0 | 0 | 0 | 0 | 0 |
|  | Wild Juverile | 0 | 0 | 0 | 0 | 0 |
|  | Total | 12,000 | 0 | 120,000 | 0 | 132,000 |
| Iceland | Hatchery Adult | 0 | 4 | 0 | 0 | 4 |
|  | Hatchery Juvenile | 40,662 |  | 0 | 0 | 40,662 |
|  | Wild Adult |  | 53 | 0 | 0 | 53 |
|  | Wild Juverile | 4,259 |  | 0 | 0 | 4,259 |
|  | Total | 44,921 | 57 | 0 | 0 | 44,978 |
| Ireland | Hatchery Adult | 0 | 0 | 0 | 0 | 0 |
|  | Hatchery Juverile | 196,293 | 0 | 0 | 0 | 196,293 |
|  | Wild Adult | 0 | 0 | 0 | 0 | 0 |
|  | Wild Juverile | 92 | 0 | 0 | 0 | 92 |
|  | Total | 196,385 | 0 | 0 | 0 | 196,385 |
| Norway | Hatchery Adult | 0 | 0 | 0 | 0 | 0 |
|  | Hatchery Juvenile | 91,878 | 13,296 | 91,000 | 0 | 196,174 |
|  | Wild Adult | 0 | 739 | 0 | 0 | 739 |
|  | Wild Juverile | 0 | 2,286 | 0 | 2,746 | 5,032 |
|  | Total | 91,878 | 16,321 | 91,000 | 2,746 | 201,945 |
| Russia | Hatchery Adult | 0 | 0 | 0 | 0 | 0 |
|  | Hatchery Juvenile | 0 | 0 | 1,022,514 | 0 | 1,022,514 |
|  | Wild Adult | 0 | 2,282 | 0 | 0 | 2,282 |
|  | Wild Juverile | 0 | 0 | 0 | 0 | 0 |
|  | Total | 0 | 2,282 | 1,022,514 | 0 | 1,024,796 |
| Sweden | Hatchery Adult | 0 | 0 | 0 | 0 | 0 |
|  | Hatchery Juvenile | 0 | 3000 | 180,343 | 0 | 183,343 |
|  | Wild Adult | 0 | 0 | 0 | 0 | 0 |
|  | Wild Juverile | 0 | 500 | 0 | 0 | 500 |
|  | Total | 0 | 3,500 | 180,343 | 0 | 183,843 |
| UK (England \& | Hatchery Adult | 0 | 0 | 0 | 0 | 0 |
| Wales) | Hatchery Juverile | 0 | 0 | 211,464 | 0 | 211,464 |
|  | Wild Adult | 0 | 250 | 0 | 0 | 250 |
|  | Wild Juverile | 3,994 | 0 | 8,820 | 891 | 13,705 |
|  | Total | 3,994 | 250 | 220,284 | 891 | 225,419 |
| UK (N. Ireland) | Hatchery Adult | 0 | 0 | 0 | 0 | 0 |
|  | Hatchery Juverile | 15,671 | 0 | 30,442 | 0 | 46,113 |
|  | Wild Adult | 0 | 0 | 0 | 0 | 0 |
|  | Wild Juverile |  | 0 | 0 | 0 | 0 |
|  | Total | 15,671 | 0 | 30,442 | 0 | 46,113 |
| UK (Scotland) | Hatchery Adult | 0 | 0 | 0 | 0 | 0 |
|  | Hatchery Juverile | 0 | 0 | 49,163 | 786 | 49,949 |
|  | Wild Adult | 0 | 626 | 42 | 166 | 834 |
|  | Wild Juverile | 2990 | 0 | 9 | 2,411 | 5,410 |
|  | Total | 2,990 | 626 | 49,214 | 3,363 | 56,193 |
| USA | Hatchery Adult | 0 | 1,695 | 0 | 2,057 | 3,752 |
|  | Hatchery Juvenile | 0 | 0 | 265,458 | 665 | 266,123 |
|  | Wild Adult | 0 | 0 | 10 | 0 | 10 |
|  | Wild Juverile | 0 | 0 | 0 | 81 | 81 |
|  | Total | 0 | 1,695 | 265,468 | 2,803 | 269,966 |
| All Countries | Hatchery Adult | 0 | 3,863 | 0 | 4,006 | 7,869 |
|  | Hatchery Juverile | 475,004 | 17,588 | 3,126,217 | 1,490 | 3,620,299 |
|  | Wild Adult | 31 | 6,247 | 52 | 470 | 6,800 |
|  | Wild Juverile | 11,931 | 17,884 | 19,490 | 9,160 | 58,465 |
|  | Total | 486,966 | 45,582 | 3,145,759 | 15,126 | 3,693,433 |

${ }_{2}^{1}$ Includes other internal tags (PIT, ultrasonic, radio, DST, etc.)
${ }_{3}^{2}$ May include hatchery fish.
${ }^{3}$ Includes external dye mark.

Table 2.8.2.1. Overview of current DCF and future data needs for Atlantic Salmon assessment/ scientific advice.

For more information about this table, see example of Table 4.2.3.1 of WKESDCF report.

| Type of data | Collected under DCF | Available to WG | Reviewed and evaluated by WG | Used in current assessment models | Future plans | Notes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| How to be filled | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially used | Keep as current DCF/ Improve sampling intensity/ No need to be collected/ (other free text) | Free text |
| Fleet capacity | No ** | No * | No | No | No need to be collected | See 'Fishing gear and effort' |
| Fuel consumption | No ** | No * | No | No | No need to be collected | Many salmon fisheries use unpowered vessels |
| Fishing gear and effort | Partially ** | Partially | Partially | Partially, but information requested by NASCO | Use for estimation of exploitation rates. <br> Improve coverage and sampling intensity in DC-MAP | Data required for all relevant areas/fisheries |
| Landings | Partially ** | Yes | Yes | Yes | Improve coverage in DC-MAP | Data required on: catch in numbers and weights for recreational and commercial fisheries in rivers, estuaries and coastal waters. |


| Type of data | Collected under DCF | Available to WG | Reviewed and evaluated by WG | Used in current assessment models | Future plans | Notes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| How to be filled | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially used | Keep as current DCF/ Improve sampling intensity/ No need to be collected/ (other free text) | Free text |
| Discards | No ** | No * | No | No | No need to be collected | Not relevant to salmon except (historically) in Faroes fishery. <br> NB: 'catch and release' fish are deliberately caught and so not classed as discards. |
| Recreational fisheries | Partially ** | Yes | Yes | Yes | Improve coverage in DC-MAP | Extent of DCF coverage unclear. <br> Complete catch data needed for all recreational fisheries (see 'Landings') |
| Catch \& Release | No ** | Partially | Partially | No - but data requested by NASCO | Include collection in DC-MAP | Data on numbers of fish caught and released required for all recreational fisheries |
| cpue dataseries | Partially ** | Partially | Partially | Partially | Improve sampling intensity in DC-MAP | Data used to generate national inputs to models |


| Type of data | Collected under DCF | Available <br> to WG | Reviewed and evaluated by WG | Used in current assessment models | Future plans | Notes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| How to be filled | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially used | Keep as current DCF/ Improve sampling intensity/ No need to be collected/ (other free text) | Free text |
| Age composition | Partially ** <br> Some ageing based on fish lengths or weights | Yes | Yes | Yes | Improve coverage and sampling intensity in DC-MAP | Extent of DCF coverage unclear; sampling intensities in other fisheries inappropriate to salmon |
| Wild/reared origin (scale reading) | No ** | Partiallyfrom other sources | Partially | Partially used information on farmed fish is requested by NASCO | Improve sampling intensity in DC-MAP | Extent of DCF coverage unclear |
| Length and weight-at-age | Partially ** | Partially | Yes | Yes - but some ageing based on fish lengths or weights | Improve sampling coverage in DC-MAP | DCF does not cover all relevant areas/fisheries; sampling intensities inappropriate to salmon |
| Sex ratio | No ** | Yesfrom other sources | Partially | Yes | Modify sampling intensity in DC-MAP | Estimates required at national/regional level every five years |


| Type of data | Collected under DCF | Available <br> to WG | Reviewed and evaluated by WG | Used in current assessment models | Future plans | Notes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| How to be filled | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially used | Keep as current DCF/ Improve sampling intensity/ No need to be collected/ (other free text) | Free text |
| Maturity | Not known | No * | No | No | No need to be collected - all returning adults are mature | DCF requires collection but extent of coverage unclear; data not required for assessments |
| Fecundity | No ** | Yes | Partially | Yes | Include collection in DC-MAP | Estimates required at national/regional level every 5 years |
| Data processing industry | No ** | No ** | No | No | No need to be collected | Requirement not clear |
| Juvenile surveys (Electrofishing ) | Partially **- <br> but not requested for Atlantic salmon in DCF | Yes | Partially | Partially | Include collection in DC-MAP | Data used to develop reference points and confirm stock status. <br> Also required for assessments under WFD |


| Type of data | Collected under DCF | Available <br> to WG | Reviewed and evaluated by WG | Used in current assessment models | Future plans | Notes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| How to be filled | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially used | Keep as current DCF/ Improve sampling intensity/ No need to be collected/ (other free text) | Free text |
| Adult census data (Counters, fish ladders, etc.) | Partially **- <br> but not requested for Atlantic salmon in DCF | Yes | Partially | Yes | Include collection in DC-MAP | Counts required for ~one river in 30. Data required to provide exploitation rates for assessments |
| Index river data <br> (Smolt \& adult trapping; tagging programmes; etc.) | Partially **- <br> but not requested for Atlantic salmon in DCF | Yes | Partially | Yes | Include collection in DC-MAP | Index rivers are identified by ICES. <br> Data used to develop reference points and inputs to assessment models |
| Genetic data (for mixed-stock analysis) | No ** | Partially | Partially for some mixedstock fisheries | Not currently | Include collection in DC-MAP - sampling in mixed-stock fisheries every 5 years | Genetic analysis is now advised to provide more reliable stock composition in mixedstock fisheries |
| Economic data | Not known | No * | No | No - but data are of use to NASCO |  | Collection of economic data would be useful to managers |


| Type of data | Collected under DCF | Available <br> to WG | Reviewed and evaluated by WG | Used in current assessment models | Future plans | Notes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| How to be filled | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially | Yes/ <br> No/ <br> Partially used | Keep as current DCF/ Improve sampling intensity/ No need to be collected/ (other free text) | Free text |
| Aquaculture data | Not known | Partially marine farm production collected | Yes | No - but information on farm production is requested by NASCO |  | Currently not required for freshwater |

Add other data type to the cells with a light blue shading, if needed.

* Not asked for by the ICES WGNAS.
${ }^{* *}$ ) Not mandatory for some or all areas/stocks/fisheries under the current DCF.


Figure 2.1.1.1. Total reported nominal catch of salmon (tonnes round fresh weight) in four North Atlantic regions, 1960-2012


Figure 2.1.1.2. Nominal catch (tonnes) taken in coastal, estuarine and riverine fisheries by country. The way in which the nominal catch is partitioned among categories varies between countries, particularly for estuarine and coastal fisheries - see text for details. Note also that the y-axes scales vary.




Figure 2.1.1.3. Nominal catch taken in coastal, estuarine and riverine fisheries for the NAC and NEAC northern and southern areas over available time-series. The way in which the nominal catch is partitioned among categories varies between countries, particularly for estuarine and coastal fisheries; see text for details. Note also that the $y$-axes scales vary.


Figure 2.1.3.1. Nominal North Atlantic salmon catch and unreported catch in NASCO Areas, 19872012.


Figure 2.2.1.1. Worldwide farmed Atlantic salmon production, 1980-2012.


Figure 2.2.2.1. Production of ranched salmon (tonnes round fresh weight) in the North Atlantic, 19802012.


Figure 2.3.1.1. The Penobscot River watershed and major tributaries divided into 15 production units. Locations of the 15 hydroelectric dams included in the Dam Impact Analysis model are denoted by dashes and the name of each dam. The map inset is the Penobscot River watershed within the state of Maine, USA.


Figure 2.3.1.2. Schematic of the processes detailed within the DIA model. Rounded rectangles indicate life cycle stages, ovals indicate additions to the population, and rectangles indicate subtractions from the population. Dashed rectangles are neither additions to nor subtractions from the population, but represent dynamics incorporated into the model. All model runs simulated ten five-year generations ( 50 years) and consisted of 5000 iterations.


Figure 2.3.1.3. Median number of two sea-winter females (top) across all Penobscot River production units in generations 1-10 for scenarios with recent conditions (i.e. Base Case) and increased marine and freshwater survival rates (i.e. Recovery). Median number of two sea-winter females (bottom) across all Penobscot River production units in generations 1-10 for scenarios with all dams turned on, implementation of the PRRP (i.e. Penobscot River Restoration Project: removal of three lower river mainstem dams), all dams turned off, mainstem dams turned off and tributary dams turned on, and mainstem dams turned on and tributary dams turned off. Freshwater and marine survival rates were set at recent (i.e. Base Case) levels in all scenarios.


Figure 2.3.2.1. Standardized abundance (a) and productivity (b) of Atlantic salmon population complexes in six regions of North America: United States (US), Scotia-Fundy (SF), Gulf of St Lawrence (GF), Quebec (QC), Newfoundland (NF), and Labrador (LB).


Figure 2.3.2.2. Chronological clustering of common trends in abundance and productivity of Atlantic salmon populations (identified from dynamic factor analysis) detects key change-points and distinguishes unique periods. The number of periods marked was determined by a broken stick model.


Figure 2.3.2.3. Synthesis of pairwise correlation results that shows direct (red) and indirect (purple) influences of climate, physical, and biological factors on Atlantic salmon population trends. The relative strength of the correlation is indicated by the width of the arrows (strong $=$ thick, weaker $=$ thin).


Figure 2.3.3.1. Proportion by mass (g) of primary prey items in the stomachs of Atlantic salmon sampled in research surveys from 1965-1971 and inshore fisheries from 2006-2010 (SALSEA NA). * Data from 2010 are incomplete and data from 2011 are currently being processed.


Figure 2.3.3.2. Standardized energy content ( $\mathbf{k J} / \mathrm{g}$ fish weight) of stomach contents from Atlantic salmon sampled along West Greenland from 2006-2010 (SALSEA WG). Several sources were used to obtain the energy equivalents of various prey items (see text for details).


Figure 2.3.3.3. North American Atlantic salmon adult returns (ICES, 2012) and mean capelin length (data obtained from DFO, 2008) over time.


Figure 2.3.4.1. Geolocation positions (as determined by the methods detailed in Chittenden et al., 2011) and environmental conditions experienced by an Atlantic salmon tagged at West Greenland in September 2010. The tag popped off as programmed on April 1st, approximately seven months after tagging (black circle indicated the release location). Swimming depth was constrained when the fish was over the shelf until eventually migrating into the Labrador Sea, at which time depths in excess of 750 m were achieved. The thermal habitat occupied decreased from approximately $5^{\circ} \mathrm{C}$ to less than $0^{\circ} \mathrm{C}$, but the fish again occupied $5^{\circ} \mathrm{C}$ water upon entering the Labrador Sea.


Figure 2.3.4.2. Temperature and depth profiles of Atlantic salmon potentially predated by a) an unknown predator, b) a Greenland shark and c) one scavenged by an Atlantic halibut in the Labrador Sea in 2010 (a) and 2011 (b and c).

hierarchical
Head of tide


Exit to Gulf of St. Lawrence


Exit to Strait of Belle Isle


Figure 2.3.4.3.1. Directed Acyclic Graph (DAG) of the state-space implementation of the Cormac-JollySeber model as an annual model (panel on the left) and the exchangeability assumptions for the probability of detections in the hierarchical Bayesian model (panel on the right).


Figure 2.3.4.3.2. Number of smolts tagged and released from the Miramichi, Restigouche, and Cascapedia rivers, and subsequently detected at the head of tide, exit of bays, and Strait of Belle Isle arrays in 2007 to 2012.


Figure 2.3.4.3.3. Posterior distributions of the annual and river origin probabilities of detection at the head of tide arrays (upper panels), the bay exit arrays (middle panel) and the Strait of Belle Isle array (lower panel). The left panels are for the annual model and the right panels are for the hierarchical model. The red ellipses in the hierarchical panels identify the posterior distributions of the hyperdistribution for the detection probabilities.

| Prior assumptions for the probability of detection $(p)$ at the Strait of Belle Isle array |  |  |  |
| :--- | :---: | :---: | :---: |
|  | $\mathrm{P}(p) \sim \operatorname{Beta}(\alpha, \beta)$ | $\mathrm{E}(p)$ | $\mathrm{CV}(p)$ |
| Uniformative | $\alpha=1, \beta=1$ | 0.5 | 0.58 |
| Informative | $\alpha=9, \beta=1$ | 0.9 | 0.10 |
|  | $\alpha=7, \beta=3$ | 0.7 | 0.20 |
|  | $\alpha=50, \beta=50$ | 0.5 | 0.10 |
|  | $\alpha=5, \beta=5$ | 0.5 | 0.30 |



Figure 2.3.4.3.4. Prior assumptions for the probability of detection at the Strait of Belle Isle array (upper table) and the associated posterior distributions for the probabilities of detection and the probabilities of survival at the previous stages of the migration (lower panels). The example shown is for the smolts. [HoT = Head of Tide; GSL = Gulf of St Lawrence; SoBI = Strait of Belle Isle].


Figure 2.3.5.1. The mean number of fry dispersing each sampling day ( $24-\mathrm{h}$ period) from the five control incubators and from the five incubators exposed to artificial night light in 2011, vertical bars show $\pm 1$ standard error.

## Street lamp:12 lx

a


Control: 0.1 lx
b


Figure 2.3.5.2. Circular plots of the diel patterns of fry dispersal in 2011in relation to hours following the onset of dusk. Data are presented for each incubator (1-5) for a) 12 lx artificial night lighting and b) 0.1 lx control conditions. Also indicated are the mean dispersal times (with $95 \%$ confidence limits) for each treatment and the onset of dawn.


Figure 2.3.5.3. Circular plot of the time of migration of individual smolts permanently leaving the Brandy Stream, in relation to sunset, with the lamp either off ( $\bullet$ ) or on ( $\circ$ ) (i.e. under street lit conditions) in (a) 2009 and (b) 2008/2009 (combined). The lines are the mean vectors (for both the light on and light off groups) and their direction indicates the mean time of downstream movement (with $\mathbf{9 5 \%}$ confidence limits).


Figure 2.3.7.1.1. Sensitivity analysis to changes in egg-to-smolt survival. (a) Different egg-to-smolt survival functions tested (bold) and associated Recruitment/Stock ratio (faint). Solid line: average egg-to-smolt survival fixed at $0.7 \%$; dotted and dashed-dotted lines: BH density-dependent function with gradient of uncertainty in density-dependence with $\alpha$ equal to $2.8 \%$ and $1.4 \%$, respectively, and $\beta$ equal to $3.010^{-9}$ and $9.910^{-10}$, respectively. (b)-(c) Time-series of posterior median estimates of egg-tosmolt survival rate (b) and smolt-to-PFA survival rate (c); line patterns correspond to egg-to-smolt survival in (a).


Figure 2.3.7.1.2. Sensitivity to the hypothesis for the PFA-to-return phase. (a) Time-series of posterior median estimates of the probability of maturing in the first year at sea obtained under the constant $\mathrm{M}_{2 \mathrm{sw}}$ / variable proportion maturing (p.mat) hypothesis (dotted line) and under the variable $\mathrm{M}_{2 \mathrm{sw}}$ / constant p.mat hypothesis (bold line). (b) Total natural mortality (\%) of 1SW fish (dashed-dotted) and 2SW fish under the constant $\mathrm{M}_{2 \mathrm{sw}}$ / variable p.mat hypothesis (dotted line) and variable $\mathrm{M}_{2 \mathrm{sw}}$ / constant p.mat hypothesis (bold line).


Figure 2.3.7.2.1. Boxplots : Marginal posterior distribution of: (a) the density-dependent mortality rate as a function of the latitude and longitude, and (b) the density-dependent mortality rate for 20 index rivers. Shaded areas correspond to the posterior predictive distribution of the parameters. Solid lines: posterior median; light shaded areas are $50 \%$ posterior probability intervals; and dark shaded areas 95\% posterior probability intervals.


Figure 2.3.8.1. Percentage of fish sampled, by relative abundance category of sea lice, by month from salmon sampled in the Miramichi River, 2005 to 2011. The numbers in parentheses below the month labels are the number of fish sampled.

## 3 Northeast Atlantic Commission area

### 3.1 NASCO has requested ICES to describe the key events of the 2012 fisheries

### 3.1.1 Fishing at Faroes in 2011/2012

No fishery for salmon has been prosecuted since 2000.

### 3.1.2 Key events in NEAC homewater fisheries in 2012

The new coastal fishery by Sami communities of the Murmansk region of the Russian Federation was started in 2010 in the White Sea, where it has never been recorded in the past. The fishery continued in the coastal areas of the White Sea in 2011 and 2012. There were no other new key events reported by NEAC countries in 2012.

### 3.1.3 Gear and effort

No significant changes in gear type used were reported in 2012, however, changes in effort were recorded. The number of gear units licensed or authorized in several of the NEAC area countries provides a partial measure of effort (Table 3.1.3.1), but does not take into account other restrictions, for example, closed seasons. In addition, these data do not indicate the number of licences that were actively utilized, or the time each licensee fished.

Trends in effort are shown in Figures 3.1.3.1 and 3.1.3.2 for the Northern and Southern NEAC countries respectively. In the Northern NEAC area, driftnet effort in Norway accounted for the majority of the effort expended in the early part of the time-series. However, this fishery closed in 1989, reducing the overall effort substantially.

The numbers of gear units in UK (England \& Wales) and in UK (Scotland) (Table 3.1.3.1) were among the lowest reported in the time-series. In Norway, the numbers of bag nets and bendnets have decreased for the past 15-20 years and 2012 numbers are the lowest in the time-series. The numbers of driftnets, draftnets, bag nets and boxes for UK (N. Ireland) for 2012 were the lowest in the time-series with only four units licensed.

Rod effort trends, where available, have varied for different areas across the time-series (Table 3.1.3.1). In the Northern NEAC area, the number of anglers and fishing days in Finland has shown an increase throughout the time period, but in 2012 remained close to the previous year and to the 5-year mean. In the Southern NEAC area, rod licence numbers have generally increased since 2001 in UK (England \& Wales), although rod licence numbers in 2012 were lower than in 2011 and lower than the 5-year average. In Ireland, there was an apparent increase in the early 1990s due to the introduction of one day licences, after which licence numbers remained stable for over a decade, before decreasing from 2002 due to fishery closures. In France the effort has been fairly stable over the last ten years but showed a slight increase for the past two years.

### 3.1.4 Catches

NEAC area catches are presented in Table 3.1.4.1. The provisional declared catch in the NEAC area in 2012 was 1240 t , 179 t below the updated catch for $2011(1419 \mathrm{t})$ and $11 \%$ and $30 \%$ below the previous 5 -year and 10-year averages respectively.

The provisional total nominal catch in Northern NEAC for 2012 ( 939 t ) was 71 t below the updated catch for 2011 ( 1009 t ) and $8 \%$ and $17 \%$ below the previous five and ten year averages respectively. Catches in 2012 were below long-term averages in most Northern NEAC countries except Sweden and Finland. The catch in Iceland in 2012 was almost half the catch in 2011 and was the smallest in the time-series.

In the Southern NEAC area the provisional total nominal catch for 2012 (301 t) fell by 109 t from 2011 and was $17 \%$ and $52 \%$ below the previous 5 -year and 10-year averages respectively. Catches in 2012 were below long-term averages in most Southern NEAC countries. There was a noticeable decrease (79 t) in the catch taken in UK (England \& Wales) in 2012 compared to 2011.

Figure 3.1.4.1 shows the trends in nominal catches of salmon in the Southern and Northern NEAC areas from 1971 until 2012. The catch in the Southern area has declined over the period from about 4500 t in 1972 to 1975 to below 1000 t since 2003, and was between 250-450 t over the last five years. The catch showed marked declines in 1976 and in 1989 to 1991 and continues to show a steady decline over the last ten years. The catch in the Northern area also indicated an overall decline over the time-series, although this decrease was less distinct than the reductions noted in the Southern area. The catch in the Northern area varied between 2000 t and 2800 t from 1971 to 1988 , fell to a low of 962 t in 1997, and then increased to over 1600 t in 2001. Catch in the Northern area has exhibited a downward trend since and is now below 1000 t . Thus, the catch in the Southern area, which comprised around two-thirds of the total NEAC catch in the early 1970s, has been lower than that in the Northern area since 1999.

### 3.1.5 Catch per unit of effort (cpue)

Cpue is a measure that can be influenced by various factors such as fishing conditions, perceived likelihood of success and experience. It is assumed that the cpue of net fisheries is a more stable indicator of the general status of salmon stocks than rod cpue, with the latter generally assumed to be more greatly affected by varying local factors such as weather conditions, management measures and angler experience. Both cpue of net fisheries and rod cpue may also be affected by measures taken to reduce fishing effort, for example changes in regulations affecting gear. If changes in one or more factors occur, a pattern in cpue may not be immediately evident, particularly over larger areas. It is, however, expected that for a relatively stable effort, cpue can reflect changes in the status of stocks and stock size. Cpue may be affected by increasing rates of catch and release in rod fisheries.

The cpue data are presented in Tables 3.1.5.1-3.1.5.6. The cpue data for rod fisheries have been derived by relating the catch to rod days or angler season. Cpue for net fisheries were calculated as catch per licence-day, trap month or crew month.

In the Southern NEAC area, cpue has generally decreased in UK (England \& Wales) and UK (Scotland) net fisheries (Figure 3.1.5.1). Cpue values for net fisheries in these coun-
tries were mostly lower figures compared to 2011 and also to the previous 5-year averages (Table 3.1.5.3 and 3.1.5.5). The cpue values for rod fisheries in UK (England \& Wales) and France in 2012 decreased from 2011 and were lower or close to the previous 5-year mean (Table 3.1.5.1 and Table 3.1.5.4). In UK (N. Ireland), the River Bush rod fishery Cpue showed an increase from 2011 but was lower than the 5 -year average (Table 3.1.5.1). Overall, there is little change in cpue levels despite major reductions in fishing effort.

In the Northern NEAC area, there has been an increasing trend in cpue figures for the Russian rod fisheries in the White Sea rivers (Figure 3.1.5.1). This trend is pronounced in the River Ponoi, the only White Sea river with reported cpue for the whole dataseries (Table 3.1.5.2). The cpue for the commercial coastal and in-river net fisheries in the Archangelsk area, Russia, showed an increase from 2011 and higher values than the 5-year average (Table 3.1.5.2). A slight decreasing trend was noted for rod fisheries in Finland (River Teno and River Naatamo) over the time-series (Figure 3.1.5.1), although in 2012, both rivers showed an increase from 2011 and to the 5 -year average (Table 3.1.5.1). An increasing trend was observed for the Norwegian net fisheries (Figure 3.1.5.1 and Table 3.1.5.6).

### 3.1.6 Age composition of catches

The percentage of 1SW salmon in NEAC catches is presented in Table 3.1.6.1 and in Figures 3.1.6.1 (Northern NEAC) and 3.1.6.2 (Southern NEAC). The overall percentage of 1SW fish in the Northern NEAC area catch remained reasonably consistent in the period 1987 to 2000 (range $61 \%$ to $72 \%$ ), but has fallen in more recent years (range $49 \%$ to $69 \%$ ), when greater variability among countries has also been evident. In the Northern NEAC area, the percentage of 1SW fish in the catch in 2012 ( $55 \%$ ) was higher than in 2011 (49\%), but equal to the previous 5-year average and lower than the previous 10-year average ( $58 \%$ ). On average, 1 SW fish comprise a higher percentage of the catch in Iceland, Finland and Russia than in the other Northern NEAC countries, with the percentage of 1SW fish in Norway and Sweden remaining the lowest among the Northern NEAC countries (Figure 3.1.6.1). The percentage of 1SW fish in catches in Iceland, Finland, Russia and Norway were higher in 2012 than in 2011, while in Sweden, the percentage of 1SW fish in catches was lower in 2012 than in 2011 (Figure 3.1.6.1).

In the Southern NEAC area, the percentage of 1SW fish in the catch in 2012 (49\%) was higher than in 2011 ( $45 \%$ ), but lower than the previous 5-year and 10-year averages $(56 \%$ and $58 \%$ respectively). The percentage of 1 SW salmon in the Southern NEAC area remains reasonably consistent over the time-series (range $45 \%$ to $65 \%$ ), although with considerable variability among individual countries (Figure 3.1.6.2). 1SW fish typically comprise a larger proportion of the catch in UK (England \& Wales) than in the other Southern NEAC countries for which data are available, although this proportion declined in 2011 and 2012.

The percentage of 1SW salmon in 2012 was among the lowest in the time-series for both the Northern and Southern NEAC areas ( $55 \%$ and $49 \%$ respectively). It should be noted that the data presented in Figures 3.1.6.1 and 3.1.6.2 may be influenced by various management measures and represent variation in age composition among years rather than within cohorts.

### 3.1.7 Farmed and ranched salmon in catches

The contribution of farmed and ranched salmon to national catches in the NEAC area in 2012 was again generally low in most countries, with the exception of Norway, Iceland and Sweden, and is similar to the values that have been reported in previous years. The occurrence of such fish is usually ignored in assessments of the status of national stocks (Section 3.3).

The estimated proportion of farmed salmon in Norwegian angling catches in 2012 was among the lowest on record (5\%), whereas the proportion in samples taken from Norwegian rivers in autumn was at a similar level to previous years (12\%). In a Kolarctic sampling programme (Section 2.3.10) (19 489 scale samples in 2011-2012, from 58 seafishers; Niemelä et al., unpublished) from the northern part of Norway (Northern-Nordland, Troms and Finnmark counties), 11\% of the salmon collected between early May and late September were escaped farmed salmon. Escaped salmon occurred within the Kolarctic area during the entire summer (May: $6 \%$; June: $22 \%$; July: $41 \%$; August: $30 \%$; September: $1 \%$. Numbers and proportions of escaped salmon increased towards the end of August in all counties. Mean weights of escaped salmon were larger than the mean weights of wild salmon in the sea catches in all counties. During the official sea fishing season, escaped salmon comprised $40 \%, 20 \%$ and $5 \%$ of the reported salmon catches in NorthernNordland, Troms and Finnmark respectively. In a sample of 138 individuals of Atlantic salmon captured on the west coast of the Spitsbergen Island at Svalbard in the period 2008-2010, 11 individuals ( $8 \%$ ) were classified as escaped farmed salmon based on genetic analysis and scale characteristics. The size of these escaped fish and their growth while at sea were similar to those of wild fish captured at the same time in the same area. This suggests that escaped farmed salmon may grow, migrate, and disperse throughout the ocean like their wild counterparts, and survive to adulthood (Jensen et al., in press).

The number of farmed salmon that escaped from Norwegian farms in 2012 is reported to be 38000 fish (provisional figure), the lowest on record. An assessment of the likely effect of these fish on the output data from the PFA model has been reported previously (ICES 2001). The release of smolts for commercial ranching purposes ceased in Iceland in 1998, but ranching for rod fisheries in two Icelandic rivers continued in 2012. Icelandic catches have traditionally been split into two separate categories, wild and ranched. In 2012, 12 t were reported as ranched salmon in contrast to 53 t harvested as wild. Similarly, Swedish catches have been split into two separate categories, wild and ranched, although ranched salmon are not reported separately as they are for Iceland (Table 2.1.1.1). In 2012, 20 t were reported as ranched salmon in contrast to 10 t harvested as wild. Ranching occurs on a much smaller scale in other countries. Some of these operations are experimental and at others harvesting does not occur solely at the release site.

### 3.1.8 National origin of catches

### 3.1.8.1 Catches of Russian salmon in northern Norway

Evidence of Russian origin salmon being caught in coastal mixed-stock fisheries in northernmost Norway has been reported in previous years (e.g. ICES, 2009a). Norway has recently decreased fishing effort in coastal areas and the available information shows a decline in the number of fishing days and in the number of fishers operating in marine
waters of Finnmark County. However, there are still significant salmon fisheries operating in this coastal area exploiting Atlantic salmon of Russian origin.

The results of a pilot study initiated in 2009 by Norway, Russia and Finland showed that coastal catches taken in northern Norway in 2008 consisted of a mix of salmon from a number of rivers in both Russia and Norway, with the Russian component in Finnmark increasing from west to east (Svenning et al. 2011). On average, through the whole sampling period from May to August, and across all sampling localities, the proportion of Russian salmon in the catches was $20 \%$. Between regions and seasons, the proportion varied, reaching levels of up to $70 \%$ of Russian salmon in catches in the Varanger area, close to the Russian border, early in the season. There were also differences in regional and temporal distributions of different sea age groups. Through the season the proportion of Russian salmon decreased in all sampling areas. The results also demonstrated that bag nets and bendnets located near the coast catch fish from a larger number of stocks than nets located in the fjords (Svenning et al. 2011).

This work is continuing under the Joint Russian-Norwegian Scientific Research Programme on Living Marine Resources in 2013 (Appendix 10 of the 42nd Joint RussianNorwegian Fishery Commission) and under the Kolarctic Salmon project (Kolarctic ENPI CBC programme) (Section 2.3.11).

### 3.1.8.2 Regulation of the salmon fishery on the English northeast coast

In UK (England \& Wales), a recent genetic analysis was used to help inform decisions about the future regulation of the salmon fishery on the English northeast coast. Samples were collected from almost two thousand salmon in 2011 and analysed by applying a genetic assignment approach. This relied on a baseline of genetic information in respect of populations of fish expected to contribute to the fishery; i.e. primarily rivers in northeast England and eastern Scotland. The ability to assign fish successfully depends on the baseline containing sufficient genetic variation to allow differentiation among the rivers/regions of interest. In practice, it was not possible to assign individual fish to their specific river of origin. However, it was possible to determine the proportions of the net catch that were from northeast English and Scottish rivers as regional groups. This confirmed the previous results from the tagging studies.

### 3.1.9 Exploitation indices for NEAC stocks

Exploitation estimates have been plotted for 1SW and MSW salmon from the Northern NEAC (1983 to 2012) and Southern NEAC (1971 to 2012) areas (Figures 3.1.9.1 and 3.1.9.2).

National exploitation rates are an output of the NEAC PFA Run Reconstruction Model. These were combined as appropriate by weighting each individual country's exploitation rate to the reconstructed returns.

Data gathered prior to the 1980s represent estimates of national exploitation rates while post 1980s exploitation rates have often been subject to more robust analysis informed by projects such as the national coded wire tagging programme in Ireland. The overall rate of change of exploitation within the different countries in the NEAC area is presented as a plot of the \% change in exploitation rate per year (derived from the slope of the linear
regression between time and natural logarithm transformed exploitation rate) over the time-series (Figure 3.1.9.3).

In 2012, exploitation of northern NEAC 1SW stocks was higher than exploitation of southern NEAC 1SW stocks. The exploitation of 1SW salmon in both northern NEAC and southern NEAC areas has shown a general decline over the time-series (Figures 3.1.9.1 and 3.1.9.2), with a notable sharp decline in 2007 as a result of the closure of the Irish driftnet fisheries in the Southern NEAC area. The weighted exploitation rate of 1SW salmon in the Northern NEAC area was $40 \%$ in 2012 representing a slight decline from the previous 5 -year ( $41 \%$ ) and 10 -year ( $43 \%$ ) averages. Exploitation of 1 SW fish in the southern NEAC complex was $12 \%$ in 2012 indicating a decrease from both the previous 5 -year $(15 \%)$ and 10 -year ( $24 \%$ ) averages. The current estimates for both stock complexes are at or among the lowest in the time-series.

Exploitation of northern NEAC MSW stocks was also higher than exploitation of southern NEAC MSW stocks in 2012. The exploitation rate of MSW fish also exhibited an overall decline over the time-series in both northern NEAC and southern NEAC areas (Figures 3.1.9.1 and 3.1.9.2), with a notable sharp decline in the northern NEAC area in 2008 as a result of significant changes in the Norwegian fisheries. Exploitation of MSW salmon in the northern NEAC area was $44 \%$ in 2012, lower than the previous 5 -year ( $49 \%$ ) and 10-year ( $53 \%$ ) averages. Exploitation of MSW fish in southern NEAC was $11 \%$ in 2012 , lower than the previous 5 -year ( $13 \%$ ) and 10 -year ( $15 \%$ ) averages.

The relative rate of change of exploitation over the entire time-series indicates an overall reduction in exploitation in most northern NEAC countries for 1SW and MSW salmon (Figure 3.1.9.3). Exploitation in Finland has been relatively stable over the time period whilst the largest rate of reduction has been for 1SW salmon in Russia. The southern NEAC countries have also shown a general decrease in exploitation rate (Figure 3.1.9.3) for both 1SW and MSW components. The greatest rate of decrease shown for both 1SW and MSW fish was in UK (Scotland) whilst Iceland (SW) showed relative stability in exploitation rates for both 1SW and MSW salmon during the time-series. 1SW exploitation rates in France showed a long-term increase, although the exploitation rate has been relatively stable over the past 18 years.

### 3.1.10 Bycatch of salmon in pelagic fisheries

The Icelandic Directorate of Fisheries (IDF) started a screening programme to investigate the incidence of salmon bycatch in mackerel/herring fisheries in 2010. In that year the programme was limited to $1000-3000 \mathrm{t}$ multi-gear vessels, which fished with a midwater trawl and landed their catch in processing factories and freezing plants. In 2011 and 2012 the screening programme continued and included larger processing and factory vessels. The screening period lasted from early June to late September and was conducted by IDF inspectors. The Icelandic mackerel/herring fishery took place from northeastern to western areas with salmon caught as bycatch in all the areas. In addition to the salmon detected and sampled by the IDF personnel, Icelandic fishermen voluntarily provided salmon to samplers. The total catch screened was recorded by the IDF as well as the number of salmon provided by fishermen and the total catch of mackerel and herring caught by each of the ships returning salmon. The IDF screened catch and the catch from ships voluntarily providing salmon is regarded as total catch screened.

In 2010, 170 salmon were recovered from a total of 35403 tonnes of mackerel and herring, an average of 4.8 salmon/ 1000 t . In 2011, a total of 40804 t were screened and 249 salmon were recovered ( 6.1 salmon/1000 t). The total catch screened in 2012 was 37349 t with a bycatch of 48 salmon ( 1.3 salmon/1000 t). Four tagged salmon caught in 2010 were from Norway and one from Ireland. In 2011, one Norwegian and one Irish tagged salmon were caught. The Norwegian salmon originated from rivers in West Norway, Daleelva, Vosso (2) and Imsa. The Irish fish were from River Shannon and River Bundorraghe, West Ireland. In these three years, no Icelandic tagged salmon have been recovered. For each salmon, information has been recorded on the date and place (coordinates) of capture, along with the length, weight and sex of the fish and details of any tags recovered. From 2011, additional samples were taken for sex and maturity determination, stomachs were retained for diet analyses and tissue was taken for DNA analyses. The head of each fish was also retained. Further screening of salmon bycatch in the Icelandic pelagic fishery is planned in 2013 as well as further analyses of existing samples.
In 2012, no information related to the pelagic fishery, screening, or bycatch of salmon was reported from the Faroes. The Working Group noted that screening of salmon as bycatch in pelagic fisheries as well as collecting of biological samples will increase the knowledge of salmon in these areas.

The Working Group recommended that similar sampling should continue in order to provide further information on the bycatch of salmon in pelagic fisheries in these areas.

### 3.2 Management objectives and reference points

Management objectives for North Atlantic salmon stocks are outlined in Section 1.4. Section 1.5 describes the derivation of reference points for these stocks and stock complexes.

The current status of the NEAC stock complexes is considered here with respect to ICES guidance. Conservation limits (CLs) have been defined by ICES as the stock level that will achieve long-term average maximum sustainable yield (MSY). NASCO has adopted this definition of CLs (NASCO 1998). The CL is a limit reference point; having populations fall below these limits should be avoided with high probability. Historically, stock complexes in the NEAC area have been interpreted to be at full reproductive capacity only if the lower bound of the $95 \%$ confidence interval of the most recent spawner estimate is above the CL. In a similar manner, the status of stocks prior to the commencement of distant water fisheries has been interpreted to be at full reproductive capacity only if the lower bound of the $95 \%$ confidence interval of the most recent pre-fishery abundance (PFA) estimate is above the Spawner Escapement Reserve (SER).

Previously, ICES (2011b) assessed the status of stocks and provided advice on management of the stock complexes in the NEAC area based on the uncertainties in the estimates of spawners relative to CLs. Specifically, if the lower bound of the $95 \%$ confidence interval (i.e. $97.5 \%$ probability) of the current estimate of spawners was above the CL, then the stock was considered at full reproductive capacity. When the lower bound of the confidence limit was below the CL, but the midpoint was above, the stock was considered to be at risk of suffering reduced reproductive capacity. Finally, when the midpoint was below the CL, the stock was considered to be suffering reduced reproductive capacity.

The risk assessment framework in this current report directly evaluates the risk of meeting or exceeding the stock complex objectives. Managers can choose the risk level which
they consider appropriate. ICES considers however that to be consistent with the MSY and the precautionary approach, and given that the CLs are considered to be limit reference points and to be avoided with a high probability, then managers should choose a risk level that results in a low chance of failing to meet the CLs. ICES recommends that the probability of meeting or exceeding CLs for individual stocks should be greater than 95\% (ICES 2012c). As such, the following terminology is used in this year's report to characterize stock status at the stock complex and country levels:

- ICES considers that if the lower bound of the $90 \%$ confidence interval of the current estimate of spawners is above the CL, then the stock is at full reproductive capacity (equivalent to a probability of at least $95 \%$ of meeting the CL).
- When the lower bound of the confidence limit is below the CL, but the midpoint is above, then ICES considers the stock to be at risk of suffering reduced reproductive capacity.
- Finally, when the midpoint is below the CL, ICES considers the stock to be suffering reduced reproductive capacity.

ICES has also indicated that for the implementation of the risk framework for the provision of catch advice for the NEAC area, management objectives should be defined for each salmon management unit. Such management objectives have yet to be agreed by NASCO and are discussed further in Section 3.4.

### 3.2.1 Setting conservation limits

River-specific CLs have been derived for salmon stocks in some countries in the NEAC area (France, Ireland, UK (England \& Wales) and Norway). An interim approach has been developed for estimating national CLs for countries that cannot provide one based upon river-specific estimates. This approach is based on the establishment of pseudo stock-recruitment relationships for national salmon stocks (Potter et al., 2004).

As described previously (ICES, 2002), the NEAC-PFA run reconstruction model (Section 3.3.1) provides a means of relating estimates of the numbers of recruits to the numbers of spawners. The numbers of 1SW and MSW spawners are converted into numbers of eggs deposited using the proportion of female fish in each age class and the average number of eggs produced per female. The egg deposition in year ' $n$ ' is assumed to contribute to the recruitment in years " $n+3$ " to " $n+8$ " in proportion to the numbers of smolts produced of ages 1 to 6 years respectively. These proportions are then used to estimate the 'lagged egg deposition' contributing to the recruitment of maturing and non-maturing 1SW fish in the appropriate years. The plots of lagged eggs (stock) against the 1SW adults in the sea (recruits) are presented (Section 3.3.4) as 'pseudo stock-recruitment' relationships for each homewater country unable to provide river-specific CLs.

ICES currently define the CL for salmon as the stock size that will result in the maximum sustainable yield (MSY) in the long term. It is not straightforward to estimate this point on the stock-recruitment relationships established by the national PFA runreconstruction models, however, as the replacement line (i.e. the line on which 'stock' equals 'recruits') is not known for these relationships. This is because the stock is expressed as eggs, while the recruits are expressed as adult salmon. The Working Group has previously adopted a method for setting biological reference points from the national
pseudo stock-recruitment datasets (ICES, 2001). This model assumes that there is a critical spawning stock level below which recruitment decreases linearly towards zero and above which recruitment remains constant. The position of this critical stock level is determined by searching for the stock value that provides the line of best fit for the stock and recruitment data provided by the PFA run-reconstruction model as determined by the residual sum of squares. This point is a proxy for Slim and is therefore defined as the CL for the stock. This approach was again applied to the 2012 national stock-recruitment relationship assessment for countries where no river-specific CLs have been determined.

### 3.2.2 National conservation limits

Where river-specific estimates of CL have been derived for the country as a whole (France, Ireland, UK (England \& Wales) and Norway), these are used to provide national estimates. For countries where the development of river-specific CLs has not been completed, the method based on the PFA run-reconstruction model and described above has been used (Table 3.2.2.1). The estimated national CLs have been summed for northern and southern NEAC stock complexes (Table 3.2.2.1).

The CLs have also been used to estimate the SERs (the CL increased to take account of natural mortality between the recruitment date, 1st January, and return to homewaters) for maturing and non-maturing 1SW salmon from the northern NEAC and southern NEAC stock complexes (Table 3.2.2.1).

The Working Group considers the current CL and SER levels may be less appropriate for evaluating the historical status of stocks (e.g. pre-1985), that in many cases have been estimated with less precision.

### 3.2.3 Progress with setting river-specific conservation limits

In Finland, information has been collated to set CLs for two tributary systems of the River Teno/Tana in addition to the five Norwegian tributaries where CL attainment has been evaluated in recent years. There is also work in progress in Norway in cooperation with Finland to revise the CLs for the Teno/Tana river system based on new information on habitat distribution and revised GIS-estimates of wetted area.

In UK (Scotland), CLs and associated measures of spawning escapement and returns to homewaters will be available for salmon stocks by summer of 2013. This information will be used to inform the management of stocks objectively, consistently and in line with NASCO guidance. CLs are derived from stock and recruitment data from the River North Esk and transported to other areas based on estimates of the relative wetted area and productivity of the salmon habitat. Estimates of stock abundance are derived primarily from fishery returns. Assessment will generally be possible down to the fishery District scale, the lowest scale at which fishery data are routinely available. Districts correspond either to a single river catchment together with adjacent coast or to groups of neighbouring river catchments and associated coastline (Scottish Government, 2012). Much of the data used to derive these models has been derived from long-term monitoring undertaken by Marine Scotland Science (MSS) on the River North Esk together with additional information from a limited number of other rivers. Biological data from a wider range of stocks is required to increase the robustness of the assessments of Scottish
stocks at the District scale. Further development is also required to allow assessment at sub-catchment scales, for example to inform management of spring salmon stocks.

In UK (N. Ireland), in addition to progress previously reported (ICES, 2012a), the River Clady has been surveyed and a CL for that river stock established.

In Ireland, CLs for salmon stocks were updated in 2012 for calculation of 2013 catch advice. This was undertaken for a number of reasons:

- to update reference rivers providing stock-recruitment indices to a more Irish focused set in light of new Irish river counter data;
- to ensure that CLs are based on up-to-date, river-specific biological information;
- in light of updated river wetted areas.

A full description of the changes proposed was provided in ICES (2012a). In summary, prior to the 2012 analyses for 2013, the Bayesian Hierarchical Stock and Recruitment Analysis (BHSRA) model was developed for a set of 13 stock and recruitment dataseries from monitored salmon rivers located in the Northeast Atlantic. For the 2012 analyses for the 2013 season the index rivers were updated, to a more Irish focused series comprising 22 rivers, of which 17 are in the island of Ireland, four in UK (England \& Wales) and one in UK(Scotland). The time-series of spawner-recruits for each river was updated and the model re-run. This yields a set of predicted stock and recruitment parameters for new rivers, provided information is available on the size of the river (in this case usable habitat or wetted area is used) and on the river's latitude.

The most current biological information was used in estimating river salmon populations, in terms of the ratio of 1SW to MSW fish; the weights of each and their associated fecundities. Prior to the 2012 analyses these values were estimated and set nationally based upon best available information. For the 2012-2013 analyses, values are riverspecific where catches of salmon less than 4 kg and greater than 4 kg were each above 50 fish between 2006 and 2011; for rivers with smaller catches, national averages were applied.

The previous sums of CLs for individual rivers in Ireland indicated a requirement of 236044 1SW fish and 15334 MSW fish ( 251378 total CL). The new estimates are 211471 1SW and 46943 MSW fish ( 258415 total CL). The new output indicates a much higher MSW requirement than previously. This is due principally to the following factors:

- The breakdown of $15 W$ and MSW is more specific to a particular river and stock i.e. in general the age distribution is based on catch weights split ( $>4 \mathrm{~kg}$ is the cut-off for MSW salmon) or catches made in spring/summer (i.e. all fish caught up to the end of May are considered MSW fish).
- The new fecundity estimates are now based on contemporary analysis of field and literature studies and indicate a lower fecundity of 6000 eggs per 2 SW female compared to the previous value of 8000 eggs per 2 SW female. This in turn will result in a higher requirement of individual fish to meet the same CL than previously.
- Estimates from the new wetted area analysis include a requirement for spawners in the upstream portions of large impounded rivers. These have now been included in the assessment of CL nationally.


### 3.3 NASCO has requested ICES to describe the status of stocks

### 3.3.1 Development of the NEAC-PFA run-reconstruction model

The Working Group has developed a model to estimate the PFA of salmon from countries in the NEAC area. The original model was described by Potter et al. (2004) and modifications have been described in subsequent Working Group reports. PFA in the NEAC area is defined as the number of 1 SW recruits on January 1st in the first sea winter. The model is based on the annual catch in numbers of 1SW and MSW salmon in each country. These are raised to take account of minimum and maximum estimates of non-reported catches and exploitation rates of these two sea-age groups. These values are then raised to take account of the natural mortality between January 1st in the first sea winter and the mid-point of the period over which the respective national fisheries operate. The Working Group determined a natural mortality value of 0.03 per month to be the most appropriate (ICES, 2002) and a range 0.02 to 0.04 is used within the model. A Monte Carlo simulation (12 000 trials) is used to estimate confidence intervals on the stock estimates.

The model has previously been run using 'Crystal Ball' (CB) in Excel (Decisioneering 1996). An updated version of the model which runs in the ' $R$ ' programming language ( $R$ Development Core Team, 2007) was developed in 2011 (ICES, 2011b). The objective was to provide a more flexible platform for the further development of the model and to allow its integration with the Bayesian forecast model for the development of catch options. In 2012, the outputs of the CB and ' R ' models were compared to examine the approaches taken and validate the outputs (ICES, 2012a). The run-reconstruction analysis in 2013 was completed using the ' $R$ ' program.

### 3.3.2 National input to the NEAC-PFA run-reconstruction model

To run the NEAC PFA run-reconstruction model, countries are generally required to input the following time-series of information, beginning in 1971, for both 1SW and MSW salmon (Annex 5):

Catch in numbers;
Unreported catch levels (minimum and maximum); and
Exploitation levels (min and max).
These inputs have been described in detail previously (ICES, 2002). Modifications to these inputs are reported in the year in which they are first implemented and modifications undertaken in 2012 are provided in Section 3.3.3.

For some countries, the data are provided in two or more regional blocks. In these instances, model output is provided for the regions and is also combined to provide output data for the country as a whole. The input data for Finland consists solely of catches from the River Tana/Teno. These comprise both Finnish and Norwegian net and rod catches. The Norwegian catches from the River Tana/Teno are not included in the input data for Norway.

With the implementation of the model in ' $R$ ', it has been agreed that where the above data are themselves derived from other data sources, the raw data should be included in the model. This will permit the uncertainty in these analyses to be incorporated into the modelling approach. Currently the model input data for UK (England \& Wales) and Greenland have been modified in this way. For UK (England \& Wales) the changes address the estimation of the catches of Scottish fish in the northeast English coastal fishery, which are incorporated into the assessment for UK (Scotland). For Greenland, catch data are input in the form of harvests (reported and unreported) in weight, along with data from the West Greenland sampling programme.

For countries where fishery data were supplied disaggregated by region, the CB model previously estimated the national CL using the total national estimates of lagged eggs and PFA (see 3.2.1). However, it was considered more appropriate (ICES, 2012a) to apply this analysis to each region and then sum the resulting estimates to provide a national figure. This was implemented in the ' R ' code.

### 3.3.3 Changes to the NEAC-PFA run-reconstruction model

Provisional catch data for 2011 were updated where appropriate and the assessment extended to include data for 2012.

For Ireland, changes in the CLs used in the model are described in Section 3.2.3.
For UK (England \& Wales), exploitation rates were recalculated using days fished as the basis for estimating these data rather than licences issued as previously.

Estimates of CLs for the northern NEAC stock complexes in 2012 were similar to those reported in 2011, declining by $5.6 \%$ for 1 SW salmon stocks while increasing by $2.0 \%$ for MSW stocks. Estimates for southern NEAC stock complexes declined compared to 2011 estimates by $5.7 \%$ for 1 SW salmon stocks and increased by $14.2 \%$ for MSW stocks. Much of the change in CL estimates for the southern NEAC stock complex may be attributed to the changes in the methodology used to estimate CLs for Irish salmon stocks.

### 3.3.4 Description of national stocks as derived from the NEAC-PFA runreconstruction model

The Working Group has previously noted that the NEAC PFA model provides the best available interpretation of information on national salmon stocks. However, there remains considerable uncertainty around the derived estimates, and national representatives are continuing to improve the data inputs on the basis of new data, improved sampling and further analysis.

A limitation with a single national status of stocks analysis is that it does not capture variations in status in individual rivers or small groups of rivers, although this has been addressed, in part, by the regional splits within some countries.
The model output for each country has been displayed as a summary sheet (Figures 3.3.4.1(a-j)) comprising the following:

- PFA and SER of maturing 1SW and non-maturing 1SW salmon.
- Homewater returns and spawners ( $90 \%$ confidence intervals) and CLs for 1SW and MSW salmon.
- Exploitation rates of $15 W$ and MSW salmon in homewaters estimated from the returns and catches.
- Total catch (including unreported) of 1SW and MSW salmon.
- National pseudo stock-recruitment relationship (PFA against lagged egg deposition), used to estimate CLs in countries that cannot provide one based upon river-specific estimates (Section 3.2.1).


### 3.3.5 Trends in the abundance of NEAC stocks

Tables 3.3.5.1-3.3.5.6 summarize salmon abundance estimates for individual countries and stock complexes in the NEAC area. The PFA of maturing and non-maturing 1SW salmon and the numbers of 1SW and MSW spawners for the northern NEAC and southern NEAC stock complexes are shown in Figure 3.3.5.1.

The 5th and 95th percentiles indicated by the whiskers in each of the plots (Figure 3.3.5.1) indicate the uncertainty in the data. The Working Group recognized that the model provides an index of the current and historical status of stocks based upon fisheries data. Errors or inconsistencies in the output largely reflect uncertainties in the estimates of these data. It should also be noted that the results for the full time-series can change when the assessment is re-run from year to year as the input data are refined.

The abundances of both maturing 1SW and of non-maturing 1SW recruits for northern NEAC (Figure 3.3.5.1) show a general decline over the time period, the decline being more marked in the maturing 1SW stock. Both stock complexes have, however, been at full reproductive capacity (see Section 3.2) prior to the commencement of distant water fisheries throughout the time-series.

1SW spawners in the northern stock complexes have been at full reproductive capacity throughout the time-series. MSW spawners on the other hand, while generally remaining at full reproductive capacity, have spent limited periods either at risk of suffering reduced reproductive capacity or suffering reduced reproductive capacity.

Similarly to northern NEAC stocks, the abundances of both maturing 1SW and of nonmaturing 1SW recruits for southern NEAC (Figure 3.3.5.1) demonstrate broadly similar declining trends over the time period. Both stock complexes were at full reproductive capacity prior to the commencement of distant water fisheries throughout the early part of the time-series. Since the mid-1990s, however, the non-maturing 1SW stock has been at risk of suffering reduced reproductive capacity in approximately $50 \%$ of the assessment years. The maturing 1SW stock, on the other hand, was first assessed as being at risk of suffering reduced reproductive capacity in 2009.

The 1SW spawning stock in the Southern NEAC stock complex has been at risk of suffering reduced reproductive capacity or suffering reduced reproductive capacity for most of the time-series. In contrast, the MSW stock was at full reproductive capacity for most of the time-series until 1997. After this point, however the stock has generally been at risk of reduced reproductive capacity or suffering reduced reproductive capacity.

Based on the NEAC run reconstruction model, the status of all NEAC stock complexes, prior to the commencement of distant-water fisheries, in the latest available PFA year was considered to be at full reproductive capacity.

### 3.3.6 Compliance with river-specific conservation limits (CLs)

The status of individual rivers with regard to attainment of national CLs after homewater fisheries is shown in Table 3.3.6.1. The total number of rivers in each country and the number which can be assessed are also shown. Numerical evaluations can only be provided for seven countries where individual rivers are assessed for compliance with CLs. The compliance estimate for France is for the MSW component and data for Norway relate to 2011. Of the seven countries, the proportion of rivers assessed for compliance with CLs ranges from $0 \%$ to $88 \%$.

### 3.3.7 Survival indices for NEAC stocks

An overview of the trends of marine survival for wild and hatchery-reared smolts returning to homewaters (i.e. before homewater exploitation) is presented in Figure 3.3.7.1 and Figure 3.3.7.2. The survival indices are the percent change in return rate between fiveyear averages for the periods 2002 to 2006 and 2007 to 2011 for 1SW salmon, and 2001 to 2005 and 2006 to 2010 for 2 SW salmon. The annual survival indices for different rivers and experimental facilities are presented in Tables 3.3.7.1 and 3.3.7.2. Return rates of hatchery released fish, however, may not always be a reliable indicator of marine survival of wild fish.

The overall trend for hatchery smolts in northern and southern NEAC areas is generally indicative of a decline in marine survival. For the wild smolts this decline is also apparent for the northern NEAC areas, however for the southern NEAC areas data are more variable with some rivers showing an increase in survival whilst other rivers show a decrease. The increase in survival in the southern NEAC area is especially notable in the 2SW data. The percentage change between the means of the five-year periods varied from an $84 \%$ decline to a $111 \%$ increase in one river (Figure 3.3.7.1). However, the scale of change in some rivers is influenced by low total return numbers, where a few fish more or less returning may have a significant impact on the percent change. The survival indices for wild and reared smolts displayed a mixed picture with some rivers above and some below the previous 5 and 10-year averages (Tables 3.3.7.1 and 3.3.7.2). The return of wild 1SW salmon to the Imsa River in Norway and the Burrishoole River in Ireland was higher than both the 5-year and 10-year averages. Also the returns of 2SW wild salmon to the Rivers Frome and Tamar in UK (England \& Wales) were above the 5-year and 10-year averages. An increase in survival for hatchery reared fish was detected in Norway for 2SW salmon on the Imsa River, and on the Ranga River in Iceland for 1SW fish (Table 3.3.7.2).

Comparison of survival indices for the 2010 and 2011 smolt years show a general decrease for 2011 compared to 2010 for wild 1SW smolts in Northern and Southern NEAC areas, with the exception of the River Imsa in Norway, and the River Burrishoole in Ireland. Increased survival for 2SW returns from the 2010 smolt year compared to 2009 was also noted in most rivers that reported MSW survival in Southern NEAC for those years. Only the River Dee in UK (England \& Wales) showed a decrease in survival of 2SW relative to the year before for the 2010 smolt cohort. This could however be a result of poor sampling conditions due to consistently high flows that prevailed in UK (England \& Wales) throughout 2012 and relatively small numbers of tagged smolts in 2010. Survival indices for hatchery smolts in the northern NEAC area for the 2011 smolt year showed a decrease relative to 2010. In the southern NEAC area survival indices for hatchery smolts
decreased in the same period, except for the Irish River Burrishoole, for which the survival index was slightly higher in 2010 compared to 2009 . The only available MSW survival index for the 2009 smolt cohort, for the River Imsa in Norway (northern NEAC), showed decreased survival relative to the previous year.

Return rates for monitored rivers have been standardized to provide indices of survival for Northern and Southern 1SW and 2SW returning adult wild and hatchery salmon in the NEAC area (Figure 3.3.7.3). In summary:

- 1 SW return rates of wild smolts to the northern NEAC area (three river indices) although varying annually, have generally decreased since 1980 ( $\mathrm{p}<0.05$ ). However, a slight improvement has been noted in recent years. This declining trend is not evident for the 2 SW wild component (three river indices) and recent return rates have shown some improvement.
- Similarly, 1SW return rates of wild smolts (seven river indices) to the southern NEAC area have generally decreased since 1980 ( $\mathrm{p}<0.05$ ). Apart from return rates in 2010 (2009 smolt year), eight of the previous ten years values were the lowest in the time-series indicating a persistent period of poor marine survival. While this declining trend is not evident for the 2SW wild component (five river indices), recent returns have also been amongst the lowest in the timeseries.
- 1 SW return rates of hatchery smolts to the northern NEAC area (four river indices) although varying annually, have generally decreased since 1980 ( $\mathrm{p}<0.05$ ). However, a slight improvement has been noted in recent years. This declining trend is not evident for the 2 SW hatchery component (four river indices) and recent return rates have shown an increase.
- 1 SW return rates of hatchery smolts to the southern NEAC area (13 river indices) although varying annually, have generally decreased since 1980 ( $\mathrm{p}<0.05$ ). Although there was a slight improvement in 2012 returns ( 2011 smolt year), six of the most recent years' values are amongst the lowest in the time-series and again indicate a persistent period of poor marine survival.

Results from these analyses are broadly consistent with the information on estimated returns and spawners as derived from the PFA model (Section 3.3.5), and suggest that returns are strongly influenced by factors in the marine environment.

### 3.4 NASCO has asked ICES to further develop a risk-based framework for the provision of catch advice for the Faroese salmon fishery reporting on the implications of selecting different numbers of management units

In responding to this question, NASCO has specifically asked ICES to advise on:

- the limitations for defining management units smaller than the current NEAC stock complexes;
- the implications of applying probabilities of achieving CLs to separate management units vs. the use of simultaneous probabilities; and
- the choice of risk levels for achieving management objectives.


### 3.4.1 Background to the risk framework model

For a number of years, NASCO has asked ICES to provide catch options or alternative management advice with an assessment of risks relative to the objective of exceeding stock conservation limits for salmon in the NEAC area. ICES (2010b) described a riskbased framework that could be used to evaluate catch options for the Faroes fishery based on the method applied for the West Greenland fishery since 2003 (ICES, 2012a).

ICES (2011b) provided a more detailed evaluation of the choice of appropriate management units to be used in the risk-based framework, taking into account relevant biological and management considerations. ICES noted that breaking the stock complexes down to at least the national level was desirable because so many river stocks are exploited by the fishery. ICES also indicated that NASCO would need to agree upon the following issues before the risk-framework could be finalized:

- season to which any TAC should apply;
- share arrangement for the Faroes fishery;
- choice of management units for NEAC stocks; and
- specification of management objectives.

In the absence of feedback from NASCO, ICES (2012a) made pragmatic decisions on these questions in order to provide full catch advice for the 2012/2013 to 2014/2015 fishing seasons. The advice was provided on the basis of the four management units comprising two age groups for each of two stock complexes because it was not possible to provide stock forecasts at a more detailed (e.g. country) level at that time. However, ICES proposed that if the risk framework was run at the stock complex level then the proportion of rivers within each country meeting their CLs should also be considered when evaluating catch options.

The following sections provide a further discussion on the implications of basing the risk framework on different management units and management objectives to assist managers in agreeing on the risk framework to be employed in future.

### 3.4.2 Management units and management objectives

## Homewater fisheries

NASCO defines the basic unit for salmon management as the river 'stock', which comprises all salmon originating from a single catchment. NASCO also recommends that salmon fisheries should be managed on the basis of river and age-specific conservation limits (CL). These CLs should therefore define the minimum numbers of 1SW and MSW spawners required in each river each year, and these may be treated as separate management units (for convenience these are referred to as 'stocks' below). Fisheries should therefore be regulated to ensure that these stocks have a high probability of meeting or exceeding their CLs. The probability level chosen varies among countries but has generally been above $75 \%$. In the case of a $75 \%$ probability choice, the management objective for the fishery would be that 'there should be a greater than $75 \%$ probability of each management unit exceeding its $C L^{\prime}$. There would be no catch options for a fishery consistent with the objective of the stock exceeding the CL if the probability falls below $75 \%$.

Probabilities of attaining the CL in a given year should not be confused with the expected performance of the stock over many years. For example, if a stock annually has a probability of $75 \%$ of exceeding its CL, then over a sequence of four years, there is a $69 \%$ chance that the stock will fail to meet its CL in at least one of the four years, a $28 \%$ chance that it will fail to meet its CL in two or more years, and a $6 \%$ chance that it will fail to meet its CL in three or more years (i.e. meet the CL in only one year or fail all together). At the $95 \%$ probability level of meeting the CL, the probabilities of meeting the CL in all four years is $80 \%$ and the probability of meeting the CL in two or fewer years decreases to $2 \%$. Therefore, the choice of the risk level of meeting the CL has important implications in the risk to the stock in its year on year performance.

|  | PROBABILITY OF MEETING THE CL |  |
| :--- | :---: | :---: |
| Single-stock example: Probability over four years of <br> meeting CL | 0.75 | 0.95 |
| All four years | $31 \%$ | $80 \%$ |
| In three or fewer years | $69 \%$ | $20 \%$ |
| In two or fewer years | $28 \%$ | $2 \%$ |
| In one or fewer years | $6 \%$ | $<1 \%$ |
| In none of the four years | $1 \%$ | Near zero |

NASCO (2009) accepts that different jurisdictions may express their management objectives for salmon fisheries in different ways, and a number manage their stocks on the basis of an egg deposition conservation limit for each river. NASCO has also conceded that for severely depleted stocks, or in the absence of river and age-specific CLs, alternative management objectives might be adopted. However, these should ideally operate in a similar way and be based on the probability of attaining a reference level.

NASCO has advised that the above principles should apply equally to single-stock and mixed-stock fisheries (MSFs), with MSFs being managed to protect the weakest stocks. This means that a homewater MSF should not operate if one or more of the individual stocks (e.g. one age group from one river stock) is not expected to achieve its management objective. For any fishery to operate, the harvest or the fishery effort should be limited to ensure that the management objective is still achieved for all stocks (e.g. the probability of exceeding the CL for each stock is still greater than the values set by management, e.g. $75 \%$ or $95 \%$ ).

Achieving this overall management objective for the MSF should be guaranteed by limiting the harvest to the exploitable surplus in the weakest stock. However, this may result in a very small fishery. It would also be possible to use information on the composition of the catch in the MSF and its variability over time, to set a harvest limit that would still ensure that each stock meets its management objective.

In practice, there is a huge variation is the size of salmon river stocks, with the largest stocks being several orders of magnitude larger than the smallest. Stock status also varies considerably, with the 'healthiest' stocks exceeding their CLs by a factor of two or more. Thus a large healthy stock may mask the shortfall in a number of weaker small stocks (or vice versa). Risks to individual stocks will also be affected by the relative varia-
bility of stocks of different sizes. Thus if large stocks are more variable (i.e. have higher CVs ) relative to small stocks, it will increase the risks to the smaller stocks.

This means that as the number of stocks exploited by a fishery increases, it becomes more difficult to achieve the management objectives in all contributing stocks. This is not only because of the practical difficulties of establishing the numbers of fish from each river that are taken, but also because as the number of stocks increases it becomes less and less likely that they will all be achieving their management objectives simultaneously. Such fisheries might be managed on the basis of a single composite CL (e.g. the sum of all CLs of the contributing stocks). However, in such a case, a higher probability limit would need to be set for the combined stock in order to protect the individual stocks. In a fishery exploiting a large number of stocks (e.g. tens to hundreds) the probability of simultaneously achieving river-specific management objectives becomes very unlikely.

Comparing MSFs that exploit different numbers of stocks, the effects of a MSF depends in part on the management strategy adopted. If the overall exploitation rate (i.e. weighted mean exploitation rate for all contributing stocks) is the same in each fishery, the risks to individual stocks will be higher for the MSFs exploiting more stocks. But if the overall harvest in each fishery is kept constant, the risk to individual stocks may decrease as the number of stocks exploited in the fishery increases (effectively because the exploitation rate will decrease), although this is predicated in part on the status of the individual stocks. In practice, management strategies for MSFs are likely to fall between these extremes.

## Distant water fisheries

The distant water fishery at Faroes caught salmon originating from both northern and southern European rivers, and the fishery at West Greenland catches salmon principally from North American and southern European rivers. Both fisheries may exploit fish from well over 1000 different river stocks. If management of the fisheries at Faroes and Greenland were based on the principles described above for homewater fisheries (i.e. all contributing river stocks exceeding their river-specific CLs with a high level of probability), there would probably be no chance of a fishery being advised. This is because the probability of all potentially exploited stocks meeting their management objectives becomes highly unlikely with such a large numbers of stocks exposed to the fishery, in addition to the wide range of stock status of rivers across the North Atlantic.

Although not formally examined, even if all the over 1000 river stocks potentially contributing to these fisheries had received spawners in excess of their CLs and were producing recruitment in excess of parental stock (i.e. surplus), there is a very high probability that an important proportion of these stocks would fail to meet their CLs in any given year. In addition, productive capacity of rivers in both NAC and NEAC has been impacted by anthropogenic factors for several hundred years and many rivers are not producing recruitment at rates expected under pristine conditions. As such, the likelihood of meeting the CLs in the 1000+ rivers in the Northwest Atlantic in the past and presently is nil.

NASCO has agreed that for the management of the distant water fisheries, 'stock complexes' should be defined, which include larger numbers (100s) of rivers. ICES currently provides advice on the basis of six North American stock complexes (i.e. five Canadian
provinces and USA) and up to 19 European stock complexes (i.e. countries and regions). For the management of the West Greenland fishery, it is only necessary to consider MSW stocks, and so management decisions are based on the status of seven management units, comprising the MSW salmon in each of the six North American stock complexes and in the whole of southern Europe. For the management of the Faroes salmon fishery, management decisions take account of the status of both 1SW and MSW salmon stocks. Catch advice is currently based on the status of 1SW and MSW stocks in southern and northern Europe, making a total of four management units.

The main effect of managing on the basis of stock complexes is that an MSF can (and normally will) operate when some river stocks are not expected to achieve their management objectives. This can occur when the expected shortfall of fish in one or more stocks is compensated for by an excess in the more healthy stocks. Given the large variability in size and status of individual river stocks, this can result in the operation of the MSF when some stocks are in a severely depleted state, although the harvest from these stocks may be small.

The current risk framework for the provision of advice for the West Greenland fishery includes two mechanisms for mitigating (in part) this risk to weak stocks. First, the Greenland fishery is allocated only a proportion (currently $40 \%$ ) of the exploitable surplus in the North American stock complexes. This means that homewater fisheries, which are allocated the balance ( $60 \%$ ) of the available surplus, can be targeted at stocks that are above their CL, or the catches may be foregone to allow stock rebuilding.

In addition, the overall management objective for the provision of advice for the West Greenland fishery requires that there should be a greater than $75 \%$ probability of the stock complexes meeting or exceeding their CLs simultaneously. For the seven stock complexes used in the assessment, this is equivalent to each stock complex having approximately a $96 \%$ probability of exceeding its reference level independently (assuming all stock complexes have the same probability) (i.e. 0.96 to the power of seven is approximately equal to 0.75 ). This is consistent with ICES (2012c) advice that a $95 \%$ probability should be set for achieving the CL in each management unit. Based on the simultaneous attainment threshold only, a fishery may still be permitted if one complex has only a $75 \%$ probability of achieving its reference level as long as all others are certain ( $100 \%$ probability) of achieving theirs. It also means that if any one stock complex falls below the $75 \%$ probability limit then no fishery would be permitted.

### 3.4.3 Implications for the Faroes fishery

## Limitations for defining management units smaller than the current NEAC stock complexes

For the provision of catch advice on the West Greenland fishery, the total CL for NAC (2SW salmon only) of about 152000 fish is assessed in six management units, which means that each unit has an average CL of about 25000 salmon. In contrast, the total CLs for each of the NEAC stock complexes are:

| Northern NEAC 1SW - | 158223 |
| :--- | :--- |
| Northern NEAC MSW - | 131356 |
| Southern NEAC 1SW - | 565183 |

The NEAC stock complexes are therefore between eight and 25 times the size of the average NAC ones. There is also a wide variation in the size and status of stocks both within and among the NEAC national stock groups. ICES (2012a) has therefore recommended that the NEAC catch advice should be based on more management units than are used at present.

The Working Group noted that the use of the share allocation provides a mechanism by which risks to individual stocks may be mitigated by managers in homewaters. Since such management decisions would need to be taken at (or below) the national level, this means that it would be appropriate to disaggregate the assessment to at least the national level.

However, ICES (2012a) also noted that there are practical limitations on the extent to which the assessments can be disaggregated. The principal requirement is for information on the composition of the potential catch at Faroes by management unit. ICES proposed a method to estimate the stock composition of the Faroes catch at a national level based on tag returns and the PFA estimates. This is inevitably an approximation and the Working Group did not consider it appropriate to apply it to smaller stock complexes than countries. Genetic stock assignment studies are underway to analyse scale samples collected at Faroes (Section 2.3.6), but these are also expected to identify no more than about ten stock complexes. Other parameter values used in the assessment are currently only available for the total fishery and not smaller stock complexes.

The Working Group therefore considered that it would be informative to managers to provide the catch options tables for the four stock complexes as well as for the ten NEAC countries.

## Implications of applying probabilities of achieving CLs to separate management units vs. the use of simu/taneous probabilities

ICES tabulates the catch advice for the West Greenland fishery to show the probability of each management unit achieving its CL (or alternative reference level) and the probability that this will be achieved by all management units simultaneously. This allows managers to evaluate both individual and simultaneous attainment levels in making their management decisions. As indicated above, the probability of simultaneous attainment of management objectives in a number of separate management units is roughly equal to the product of the probabilities in each management unit. The probability threshold in each individual management unit might reasonably be set at a fixed level unless there are specific reasons for adopting an alternative (e.g. for stock rebuilding). ICES (2012a) recommended that an appropriate probability level for individual stock complexes would be $95 \%$ and this is approximately equivalent to the current management objective for the West Greenland fishery. This probability level can be applied to each management unit regardless of the number being used.

Management decisions for the West Greenland fishery have also been based on a $75 \%$ probability of simultaneous attainment of CLs. Note that for a given level of probability of achieving individual stock CLs, the probability of simultaneous attainment decreases rapidly as the number of management units increases (Figure 3.4.3.1). For the example of

20 management units (e.g. two age groups from each of ten countries), the use of the same simultaneous probability used for West Greenland (75\%) would require that the probability of individual stocks meeting the CLs be $98.6 \%$ or higher (rather than the approximately $96 \%$ value that would apply in the case of West Greenland with seven stock complexes). The use of a $95 \%$ probability level for meeting the CLs in the 20 management unit example, equates to a simultaneous attainment probability of about $36 \%$; i.e. there is a $64 \%$ chance that one or more stocks would fail to meet their CL by chance alone. Also note that the use of a $75 \%$ simultaneous objective could result in a fishery being advised when the probability of one management unit out of 20 being above its CL is as low as $75 \%$ if all the other management unit have a $100 \%$ chance of meeting the CL. This may not be an acceptable risk for managing multiple river stocks.

The Working Group considered that the probability of simultaneous attainment calculation provides useful information to managers of the risk of failing to meet CLs in one or more of the stocks in the mixed-stock fishery. However, as the management units being considered by NASCO for managing the mixed-stock fisheries at Faroes are still very large and each encompasses a large number of individual river stocks, choosing a high probability level (such as 95\%) of attaining CLs in individual stocks would be less risky to individual stocks than the use of a simultaneous attainment objective set at the value used for the West Greenland fishery.

## Choice of risk levels for achieving management objectives

On the basis of the above considerations, the Working Group agreed to provide both individual and simultaneous probabilities in the catch options tables. They also noted that management decisions should be based principally on a $95 \%$ probability of attainment of CLs in each stock complex individually. The simultaneous probability may also be used as a guide, but managers should be aware that these will generally be quite low when large numbers of management units are used.

### 3.4.4 Modelling approach for the catch options risk framework

The process for assessing each catch option within the risk framework was described by ICES (2012a). The main changes to the approach in 2013 relate to its application at a country level, although the basic model is the same. The PFA forecasts derived in the Winbugs model (See Section 3.5) are transferred to the risk framework model run in ' R '. The estimates and distributions of the PFA estimates used in the risk framework are derived by taking the first 50000 values from the Winbugs posterior forecast simulations. Parameters in the following description that are marked with an ${ }^{* \prime \prime}$ in the equations have uncertainty around them generated by means of 50000 random draws from the annual values observed from the sampling programmes conducted in the Faroes between the 1983/1984 and 1990/1991 fishing seasons. They therefore contribute to the estimation of the probability density function around the potential total harvest arising from each TAC option. When the assessment is run at a national level, the number of draws has to be limited to 25000 because of memory limitations in ' R '.

The modelling procedure involves:

- estimating the total number of $15 W$ and MSW salmon that could be killed as a result of any TAC at Faroes, including catches in homewaters;
- adjusting these to their equivalent numbers at the time of recruitment to the Faroes fishery;
- subtracting these from the PFA estimates for maturing and non-maturing 1SW salmon in the appropriate years;
- assessing the results against the SERs (i.e. the CLs adjusted to the time of recruitment to the Faroes fishery).

The TAC option (T) is first divided by the mean weight $\left(\mathrm{Wt}^{*}\right)$ of salmon caught in the Faroes fishery to give the number of fish that would be caught, and this value is converted to numbers of wild fish (Nw) by multiplying by one minus the proportion of fish-farm escapees in samples taken from the Faroes catch $\left(\mathrm{pE}^{*}\right)$ observed in historical sampling programmes. A correction factor $(\mathrm{C}=0.63)$ is applied to the proportion of fish-farm escapees to take account of reductions in the numbers of farm escapees over the past 20 years based on observations in Norwegian coastal waters:

$$
\mathrm{Nw}=\mathrm{T} / \mathrm{Wt}^{*} \times\left(1-\left(\mathrm{pE}^{*} \times \mathrm{C}\right)\right)
$$

This value is split into numbers by sea age classes (1SW and MSW) according to the proportion of each age group ( $\mathrm{pAi}^{*}$, where ' i ' is $1 S W$ or MSW) observed in historical catch sampling programmes at Faroes. In the past, there has also been a requirement to discard any fish less than 60 cm total length caught in the Faroes fishery, and $80 \%$ of these fish were estimated to die, so these mortalities are also added to the 1SW catch. Thus:
Nw1SW = Nwtotal x pA1SW* + (Nwtotal x pD* / (1-pD*) x 0.8)
and
NwMSW $=$ Nwtotal $\times$ pAMSW*
where $\mathrm{pD}^{*}$ is the proportion of the total catch that is discarded (i.e. fish $<60 \mathrm{~cm}$ total length).

Further corrections are made to the 1SW and MSW numbers to reduce the 1SW total to take account of the proportion that will not mature as 1SW fish and to add the survivors from this group to the MSW fish in the following year. For the first catch advice year the number added to the MSW total is adjusted to the TAC of the current season (i.e. zero in 2012/2013). Thus:
Nw1SW = Nw1SW x pK *
and
NwMSW = NwMSW + Nw1SW x (1-pK*)
where ' $p K$ ' is the proportion of $1 S W$ salmon that are expected to mature in the same year (0.78) derived from experimental studies conducted in the 1980s (Youngson and Webb, 1993).

The numbers in each age group are then divided among the management units by multiplying by the appropriate proportions ( pUij ), where ' i ' denotes the age groups and ' j ' denotes the management units, and each of these values is raised by the Faroes share allocation (S) to give the total potential harvest (Hij) of fish from each management unit and sea age group:
Nwij = (Nwi x pUij) / S

Finally, these values are adjusted for natural mortality so that they can be compared with the PFA forecasts and SER values from the mid-date of the fishery to the recruitment date by using an instantaneous monthly rate of mortality of 0.03 .

These harvests are then subtracted from the stock forecasts (PFAij) for the management units and sea age groups and compared with the Spawner Escapement Reserves (SERij) to evaluate attainment of the management objective. In practice, the attainment of the management objective is assessed by determining the probability that PFAij - Hij - SERij is greater than zero. The SER is the number of fish that need to be alive at the time of the Faroes fishery to meet the CL when the fish return to homewaters; this equals the CL raised by the mortality over the intervening time. CLs and SERs are currently estimated without uncertainty.

### 3.4.5 Input data for the risk framework

The analysis estimates probability of each management unit achieving its SERs (the overall abundance objective) for different catch options in the Faroes fishery (from 0 to 200 t ). The analysis assumes:

- no fishery operated in the 2012/2013 season;
- the TAC allocated to Faroes is the same in each year and is taken in full;
- homewater fisheries also take their catch allocation in full.

The assessment requires input data for the catch that would occur at the Faroes if a TAC was allocated (e.g. mean ages and weights, discard rates, etc.). In most cases the only data available to estimate these parameters come from sampling programmes conducted in commercial and research fisheries in Faroese waters in the 1980s and 1990s.

Mean weights: Mean weights of salmon caught in the commercial and research fisheries operating in Faroese waters between 1983/1984 and 1995/1996 varied between 3.06 and 5.23 kg (Table 3.4.5.1) (ICES, 1997a). However, high values were observed at the beginning of the time-series when part of the catch was taken to the north of the Faroes Exclusive Economic Zone (EEZ), and the values for the latter part of the series are based on relatively small catches in a research fishery which may not be as representative of a full commercial fishery. As a result, mean weights have been drawn randomly from the observed values of the 1985/1986 to 1990/1991 fishing seasons.

Proportion by sea age: The age composition of catches in the Faroes fishery has been estimated from samples collected in the 1983/1984 to 1994/1995 fishing seasons (Table 3.4.5.2) (ICES, 1996). The samples taken between 1991/1992 and 1994/1995 were from the research fishery and included potential discards but excluded farm escapees. As a result, values have been drawn from the observations between 1985/1986 and 1990/1991 to provide a probability distribution for this parameter. However, the age composition of the catches may be expected to be related to the mean weight. To take account of this relationship, the values of mean weight and age composition used in each sample run have been drawn from the same years.

Discards: Discard rates have been estimated from the proportions of fish less than 60 cm in catch samples between the 1982/1983 and 1994/1995 seasons (ICES, 1996) (Table 3.4.5.3); $80 \%$ of these fish were expected to die (ICES, 1986). A probability distribution for the discard rate has been estimated by sampling at random from the annual values seen over the same period as for the other parameters above.

Proportions of fish-farm escapees: The proportion of fish-farm escapees in the catches at Faroes has also been estimated from samples taken in the 1980/1981 to 1994/1995 fishing season (ICES, 1996). However, the Working Group is aware that there have been substantial changes in the production of farmed fish and in the incidence of escape events. Data were also available to the Working Group on the proportion of farm escapees in Norwegian coastal waters between 1989 and 2008; the proportion in recent years (2002-2008) was $63 \%$ of the proportion during the period 1989/1990 to 1994/1995 when the sample time-series overlap (Table 3.4.5.4). The probability distributions of proportion of farm escapees used in the risk framework has therefore been generated by multiplying the rates observed in the Faroes fishery between 1988/1989 to $1994 / 1995$ by 0.63 and then drawing sample values at random.

Proportions of catches by management unit: The origin of the stocks exploited at Faroes has previously been estimated from smolt and adult tagging studies and an approximate split between jurisdictions has been employed in the NEAC run reconstruction model (ICES, 2010b). ICES (2012a) reviewed this approach and proposed an alternative method for estimating the split of the catches. Other input parameters are shown in Table 3.4.5.5.

### 3.5 Pre-fishery abundance forecasts

The Working Group used a forecast model developed in a Bayesian framework (ICES, 2011b) to estimate PFA for all four NEAC stock complexes.

### 3.5.1 Description of the forecast model

In 2013, the Working Group ran forecast models for the southern NEAC and northern NEAC complexes. The model was run for each stock complex independently.
The PFA is modelled using the summation of lagged eggs from 1SW and MSW fish (LE) for each year $t$ and an exponential productivity parameter (a).

$$
P F A_{t}=L E_{t}^{*} \exp \left(a_{t}\right)
$$

The productivity parameter (a), is the proportionality coefficient between lagged eggs and PFA. This is forecasted one year at a time $\left(a_{t-1}\right)$ in an auto correlated random walk, using the previous year's value (a) as the mean value in a normal distribution, with a common variance for the time-series of $a$.

$$
a_{t+1}=a_{t}+\varepsilon ; \quad \varepsilon \sim N\left(0, a \cdot \sigma^{2}\right)
$$

The maturing PFA (denoted PFAm) and the non-maturing PFA (denoted PFAnm) recruitment streams are subsequently calculated from the proportion of $P F A$ maturing ( $p . P F A m$ ) for each year $t$. p.PFAm is forecast as an autocorrelated value from a normal distribution based on a logit scale, using the previous year's value as the mean and a common variance across the time-series of $p . P F A m$.
logit.p.PFAm $m_{t+1} \sim N\left(\right.$ logit.p.PFAmt , p. $\left.\sigma^{2}\right)$

## logit. p. PFAm $m_{t}=\operatorname{logit}\left(p . P F A m_{t}\right)$

Uncertainties in the lagged eggs were accounted for by assuming that the lagged eggs of 1SW and MSW fish were normally distributed with means and standard deviations derived from the Monte-Carlo run reconstruction at the scale of the stock complex. The uncertainties in the maturing and non-maturing PFA returns are derived in the Bayesian forecast models.

Catches of salmon at sea in the West Greenland fishery (as 1SW non-maturing salmon) and at Faroes (as 1SW maturing and MSW salmon) were introduced as covariates and incorporated directly within the inference and forecast structure of the model. For southern NEAC, the data were available for a 35-year time-series of lagged eggs and returns (1978 to 2012). For northern NEAC, data were available for a 22-year time-series, 1991 to 2012. The models were fitted and forecasts were derived in a consistent Bayesian framework.

For both southern and northern NEAC complexes, forecasts for maturing and nonmaturing stocks were derived for five years, from 2012 to 2016. Risks were defined each year as the posterior probability that the PFA would be above the age and stock complex specific SER levels. For illustrative purposes, risk analyses were derived based on the probability that the maturing and non-maturing PFAs would be greater than or equal to the maturing and non-maturing Spawner Escapement Reserves (SERs) under the scenario of no exploitation, for both the northern and southern complexes. These were calculated for each of the five forecast years, 2012 to 2016.

A country disaggregated version of the Bayesian NEAC inference and forecast model was run on country level data, for both southern and northern NEAC. This incorporated country specific catch proportions at Faroes, lagged eggs and returns of maturing and non-maturing components. The model was again run at the complex levels, incorporating individual country inputs of 1SW and MSW lagged eggs, 1SW and MSW returns, and SERs. Model structure and operation is as described above, incorporating country and year indexing. Linkage between countries in the model is through the common variance parameter associated with the productivity parameter (a) (the proportionality coefficient between lagged eggs and PFA), which is forecast forward and used along with the forecast proportion maturing to estimate the future maturing and non-maturing PFAs. The evolution of $a$ is independent between countries with the exception of its associated variance. Evolution of the proportion maturing ( $p . P F A m$ ) is also independent for each country, as is its variance.

### 3.5.2 Results of the NEAC Bayesian forecast models

The trends in the posterior estimates of PFA for both the southern NEAC and northern NEAC complexes closely match the PFA estimates derived from the run reconstruction model (Section 3.3.5).

For the southern NEAC stock complex, maturing and non-maturing PFAs showed a general switch in 1988-1990 from a level of around 2 million maturing and 1.5 million nonmaturing fish to lower levels, declining in the maturing component from around 1.5 million maturing in 1992-1995 to less than 1 million in 2009 (Figure 3.5.2.1). The nonmaturing component fell from around 1 million in 1990 to 1994 to a low of around 520000 in the late 2000s. The maturing component showed an increase in 2009 to 2010
which again fell in 2011, following the general slow decline from the 1990s. The nonmaturing component has followed a similar pattern in recent years, with the increase from 2008 to 2010 falling in 2011. These levels of PFA are carried forward into the five forecast years 2012 to 2016. Uncertainty increases as prediction year progresses and the forecast projects forwards, mostly notably in their upper ranges.

The productivity parameter peaked in 1985 and 1986, and reached the lowest values in 1997 and 1999 (Figure 3.5.2.1). There was a sharp drop in the productivity parameter during 1989 to 1991 and the values post-1991 are all lower than during the previous time period. Between 2003 and 2009, productivity was declining. In 2010, however, productivity rose to the 2002 to 2003 level and the highest level since the 1989 to 1991 drop, although fell again in 2011. The 2012 to 2016 values are projected from this 2011 point, with increasing uncertainty as prediction years increase.

Over the entire time-series, the maturing proportions averaged about 0.6 , being lowest in 1980 and highest in 1998 (Figure 3.5.2.1). An increasing trend in the proportion maturing (eight of 13 values below the average during 1978 to 1990 compared with four of 17 values between 1991 and 2007) is followed by a decline from 2008 which continues into 2011. The five forecast years 2012 to 2016 are projected from the 2011 value with uncertainty increasing with prediction years.

For the northern NEAC complex, peak PFA abundances were estimated in the year 2000, (ca. 1.1 million maturing and 900000 non-maturing fish) with the lowest values of the maturing series occurring between 2007 and 2009 at around 440000 fish. The lowest abundance of the non-maturing age group was estimated during 1996 to 1998 (ca. 500000 ). An increase in the year 2000 to 900000 fish was followed by a period of decline and variation with an increase from $2008(600000)$ to $2010(800000)$ (Figure 3.5.2.2). The 2012 forecasts of 466900 maturing and 675000 non-maturing fish are similar to 2011 values, with subsequent small increases predicted in the forecasts for 2013 to 2016 and increasing error bounds as forecast years progress.
Lagged eggs of both maturing and non-maturing peaked in 2007-2008 before declining to a minimum in 2010-2011 (Figure 3.5.2.2). Lagged eggs from the maturing component are estimated to continue falling during 2012 to 2016, while non-maturing are predicted to increase to above the 1996 to 2000 levels by 2015.

The proportion maturing has varied around 0.55 over the time-series, but in 2007 there was an abrupt drop in the proportion maturing to below 0.37 (Figure 3.5.2.2). Some recovery occurred in 2008 to around 0.43 ; however levels from 2009 to 2011 were around 0.40, notably below the 1991 to 2006 levels (Figure 3.5.2.2).

Productivity improved in 2009 and 2010 from the 2008-2009 low in the northern NEAC complex, becoming comparable to the pre-2004 values, but dropped slightly for 2011 (Figure 3.5.2.2). An increase was also observed in the southern complex in 2010; however this fell again in 2011 closer to 2008 and 2009 values (Figure 3.5.2.1).

Forecasts from these models for the 2012 to 2016 years for the non-maturing and maturing age groups were developed within the Bayesian model framework. Variations in the median abundance over the forecasts are related to variations in lagged eggs as the productivity parameter values are set at the level of the last year with available data. The variability in the productivity parameters increased sequentially over the forecasts.

### 3.5.3 Probability of attaining PFA above SER

Probabilities of meeting age and complex specific SERs in PFA years 2012 to 2016 do not exceed $85.3 \%$ in the southern NEAC region. For the northern NEAC stock complex, probabilities of exceeding SER are above $94.6 \%$ in all instances (Table 3.5.3.1).

### 3.5.4 Bayesian forecast models at the country level

Country specific model results of maturing and non-maturing PFA, 1SW and MSW lagged eggs, proportion of PFA maturing and productivity parameters are presented in Figures 3.5.4.1 to 3.5.4.11 for southern and northern NEAC countries. Country specific summary text-tables of the probabilities of PFAs meeting or exceeding SERs in the forecast years (2012 to 2016) are also presented in the figures. While country forecasts differ they tend to follow similar patterns to those observed in their respective stock complexes and in relation to their SERs.

Of note in these forecasts for southern NEAC countries:

- France maturing PFA is well below the SER while median non maturing PFA estimates are similar to the SER and the annual proportions maturing have high uncertainties.
- Ireland: the median estimates of maturing PFA are at or around the SER, while non-maturing PFAs are below SER. Estimates of proportion maturing are high at around 0.9.
- UK(N. Ireland): PFA maturing median estimates are at or around the SER. Estimated 25th BCIs of non-maturing PFAs are above the SER, although the probability of the non-maturing PFA achieving or exceeding the SER is above $95 \%$ in only 2012. Upper estimates are extremely variable for the forecast years of both maturing and non-maturing PFA.
- UK(England \& Wales): maturing PFA median estimates are at or around the SER and the estimated 25th BCIs of non-maturing PFAs are above the SER although the highest probability of attaining/exceeding an SER is only 0.87 (nonmaturing in 2012). The proportions maturing are approximately 0.4 although BCIs are wide.
- UK(Scotland): Patterns of maturing and non-maturing PFA in relation to their SERs and proportions maturing follow similar trends to those seen for UK(England \& Wales). Highest probability of attaining or exceeding SER is only 0.79 (for non-maturing PFA in 2012).
- Iceland (south and west): For both maturing and non-maturing PFAs the forecasts are above the respective SERs and probabilities of exceeding SERs are greater than 0.95 for maturing PFA from 2012 to 2015 and at 0.95 for nonmaturing PFA in 2012.

Of note in these forecasts for northern NEAC countries:

- Russia: For maturing PFAs the forecasts are above the SERs though the bottom of the $95 \% \mathrm{BCI}$ range falls below the SER while for the non-maturing estimates the lower 25th BCI is close to the SER. Only maturing PFA in 2012 has a 0.95 probability of exceeding SER. The proportion maturing appears to have risen
from ca. 0.4 around $2007-2010$ to ca. 0.6 in 2011, which is forecast forwards at this level to 2016.
- Finland: Both maturing and non-maturing SERs are close to the lower 25th BCIs of the PFA estimates; however a 0.95 probability of achieving SER is only seen in the maturing PFA in 2012. 1SW lagged eggs are being estimated at an all-time low between 2013 and 2016.
- Norway: There is a 0.95 probability that forecasts of non-maturing PFAs are above the SER for all forecast years (2012 and 2016). Maturing PFA achieves SER with a 0.95 probability in 2012 and 2013. MSW lagged eggs are being estimated to rise from 2013 and 2016, while the proportion maturing is being forecast forward from the 2011 level of around 0.3.
- Sweden: Maturing and non-maturing PFA estimates are above respective SERs with greater than 0.95 probability in all years for both maturing and nonmaturing PFAs. Similar to Norway MSW lagged eggs are estimated to increase from 2013 to 2016.
- Iceland (north/east regions): Maturing and non-maturing PFA forecasts are above the SERs though the probability of exceeding SER is only greater than 0.95 in 2012 and 2013 for the maturing component of the stock and 2012 for the non-maturing fish. MSW lagged eggs are estimated to increase from 2013 to 2016.


### 3.6 NASCO has asked ICES to provide catch options or alternative management advice for 2012-2015, with an assessment of risks relative to the objective of exceeding stock conservation limits and advise on the implications of these options for stock rebuilding

### 3.6.1 Catch advice for Faroes

The risk framework described in Section 3.4 has been used to evaluate catch options for the Faroes fishery in the 2013/2014, 2014/2015 and 2015/2016 fishing seasons (October to May). The assumptions and data used in the catch options assessment are described in Section 3.4.5 and the details of the forecasting methodology is described in Section 3.5.

The Working Group applied the risk framework model to the four management units previously used for the provision of catch advice (maturing and non-maturing 1SW recruits for northern and southern NEAC) and also for the two age groups in each NEAC country (i.e. 20 management units). The risk framework estimates the uncertainty in meeting defined management objectives at different catch levels (TAC options), and the catch advice tables display the probability of the total number of fish in each of the management units exceeding the combined SERs for those units.

As an example, taking the 20 t TAC option, the assessment estimates that this would result in a catch of about 5000 fish at Faroes. The great majority ( $>97.5 \%$ ) of these would be expected to be MSW fish. Once the sharing allocation is applied, and the numbers are adjusted for natural mortality to the same seasons as the PFA, this equates to about 650 maturing and 84000 non-maturing 1SW fish equivalents lost to all fisheries. The maturing 1SW component is split equally between northern NEAC and southern NEAC, while the non-maturing 1SW fish are split 0.44 to 0.54 between northern and southern areas
(with a small unaccounted proportion). These values are deducted from the PFA values which are then are compared with the following SERs:

| Northern NEAC maturing 1SW- | 201014 |
| :--- | :--- |
| Northern NEAC non-maturing 1SW- | 222888 |
| Southern NEAC maturing 1SW- | 715358 |
| Southern NEAC non-maturing 1SW- | 463566 |

The probability of the northern and southern NEAC stock complexes achieving their CLs for different catch options are shown in Table 3.6.1.1 and Figure 3.6.1.1. The probabilities with a zero TAC are the same as the values generated directly by the forecast model (Section 3.5). The catch option table indicates that the northern NEAC management units have a high probability ( $>95 \%$ ) of achieving their CLs for TACs at Faroes of up to $\sim 60 \mathrm{t}$ in the 2013/2014 season and up to $\sim 40 \mathrm{t}$ in $2014 / 2015$ to $2015 / 2016$. However, the southern NEAC stock complexes all have less than $95 \%$ probability of achieving their CLs in each year and at every TAC option. There are therefore no catch options that ensure a greater than $95 \%$ probability of each stock complex achieving its CL, and none that gives a greater than $60 \%$ probability of simultaneous attainment of CLs in all four stock complexes.

The slope of the curves in the catch option figures (Figure 3.6.1.1) is a function of the uncertainty in the estimates and the level of exploitation on the stocks resulting from a particular TAC in the Faroes fishery (Table 3.6.1.2); more uncertain data and lower exploitation rates generate flatter profiles. The flatness of the risk curves for the 1SW stocks indicates that the risk to these management units is affected very little by any harvest at Faroes, principally because the exploitation rate on these stock components in the fishery is very low (Table 3.6.1.2).

In 2012 (ICES, 2012a), the forecast model gave optimistic estimates of abundance for 2012 to 2015 because a high productivity was estimated for 2010; the model applied this value forward in forecasting the PFA in 2011 to 2015 (Section 3.5.2). In the 2013 forecasts, these productivity estimates have been reduced by the 2011 value, thus giving lower stock forecasts for 2012 to 2016.

The catch options for the country level management units are shown in Tables 3.6.1.3 and 3.6.1.4 for maturing 1 SW and non-maturing 1 SW salmon respectively from each NEAC country. The probabilities of maturing 1SW national stocks achieving their CLs in $2013 / 2014$ vary between $28 \%$ and $98 \%$ with no TAC at Faroes, and these probabilities are hardly affected by any of the TAC options. The probabilities for the two subsequent seasons are similar. The probabilities of the non-maturing 1SW national stocks achieving their CLs in 2013/2014 varies between $27 \%$ and $100 \%$ and show decreasing probabilities for increasing TAC options at Faroes. The probability of simultaneous attainment of CLs in all of the ten maturing 1 SW national stock complexes is less than $2 \%$ in every year regardless of any harvest at Faroes, and for the ten non-maturing 1SW stock complexes is less than $5 \%$ in every year.

ICES (2012a) emphasized the problem of basing the risk analysis on management units comprising large numbers of river stocks and recommended that in providing catch advice at the age and stock complex levels for Northern and Southern NEAC, consideration
should be given to the recent performance of the river stocks within individual countries. At present, insufficient data are available to the Working Group to assess performance of individual stocks in all countries in the NEAC area. In some instances CLs are in the process of being developed (UK(Scotland) and Iceland). The status of river stocks within each country in the NEAC area for which data are available with respect to the attainment of CLs before homewater fisheries is given in Table 3.6.1.5 for 2012 (except Norway where the data relate to 2011). The total number of rivers in each country and the number which can be assessed against river-specific CLs are also shown. Numerical evaluations can only be provided for three countries where individual rivers are assessed for compliance prior to homewater fisheries taking place. In two countries in northern NEAC with available information, $86 \%$ and $88 \%$ of the monitored rivers would have met their river-specific CLs before any homewater exploitation, whereas only $41 \%$ of assessed rivers met their CLs in one country in Southern NEAC (Table 3.6.1.5). So, despite the absence of a fishery at Faroes since 1999, and reduced exploitation at West Greenland on the MSW southern NEAC component, the abundance at the PFA stage in several countries in NEAC has been below the country-specific SERs.

The Working Group therefore notes that there are no catch options for the Faroes fishery that would allow all national or stock complex management units to achieve their CLs with a greater than $95 \%$ probability in any of the seasons 2013/2014 to 2015/2016. While stocks remain in a depleted state and in the absence of a fishery at Faroes, particular care should be taken to ensure that fisheries in homewaters are managed to protect stocks that are below their CLs.

### 3.6.2 Relevant factors to be considered in management

The management of a fishery should ideally be based upon the status of all river stocks exploited in the fishery. Fisheries on mixed-stocks pose particular difficulties for management, when they cannot target only stocks that are at full reproductive capacity. Management objectives would be best achieved if fisheries target stocks that are at full reproductive capacity. Fisheries in estuaries and especially rivers are more likely to meet this requirement. The Working Group also emphasized that the national stock CLs are not appropriate to the management of homewater fisheries. This is because fisheries in homewaters usually target individual or smaller groups of river stocks and can therefore be managed on the basis of their expected impact on the status of the separate stocks. Nevertheless, the Working Group agreed that the combined CLs for national stocks exploited by the distant water fisheries could be used to provide general management advice at the level of the stock complexes. As noted in previous years, the inclusion of farmed fish in the Norwegian catches could result in the stock status being overestimated (Potter and Hansen, 2001).

### 3.7 NASCO has asked ICES to update the framework of indicators used to identify any significant change in previously provided multiannual management advice

### 3.7.1 Background

In 2012 a Framework of Indicators (FWI) for the NEAC area was provided to NASCO (ICES, 2012a). The FWI was applied in January 2013, and the indicators suggested that
the PFA for non-maturing 1SW salmon from the southern NEAC area had been overestimated by the forecast model and a full reassessment was suggested for the NEAC stocks. Comparing the forecast from last year to the realized median PFA (see Section 3.3.5), all stock complexes and age groups were lower than the median of the forecasted values ( $80-94 \%$ of the forecasted value), but all were well within the $95 \%$ confidence intervals of the forecasts. This was in line with what was suggested by the FWI. The reassessment carried out in 2013 also suggested no catch options at Faroes for the coming three year period (see Section 3.6.1). Thus, carrying out the reassessment did not change the previously provided advice.

### 3.7.2 Progress in 2013

In 2012, it was agreed that the $75 \%$ confidence interval range for the mean of the indicator prediction, relative to the median forecast value, be used to compute thresholds for whether the indicator suggests a reassessment or not (ICES, 2012a). The limits should be computed at the median values of the PFA forecasts in each of the years in multiyear advice. The Working Group further agreed that if the FWI suggests that the forecasted PFA is either an underestimation or an overestimation of the realized PFA in any of the four stock complexes, then this should trigger a reassessment.

When the stocks are divided into alternative, smaller management units, potential indicators for each management unit become relatively scarce. Therefore, the Working Group recommended that the FWI be regressed against the stock complexes that they belong to. For example, MSW indicators from Norway should be regressed against PFA MSW for northern NEAC. This approach was continued in 2013. Since the indicators suggested that the forecasts were slightly higher than the realized PFA and no catch options were given in the multiyear advice for the Faroes fishery, the change indicated by the FWI did not alter the current catch advice. Therefore, in 2013, the Working Group suggests a slight change to the operation of the FWI. In the event of a closed fishery, the indicators should be compared only to the upper $75 \%$ confidence limit, and in the event of an open fishery they should be compared to both the upper and lower $75 \%$ confidence limits. This implies going from a two-sided approach to a one-sided approach in the case of a closed fishery. Had this approach been used in 2012, the reassessment would not have been suggested.

In 2013, the Working Group further updated the FWI for the NEAC area. The Working Group considered 53 possible indicator datasets, and 26 of them fulfilled the previous criteria (ICES, 2012a) for inclusion in the FWI (five for northern NEAC 1SW PFA, three for northern NEAC MSW PFA, five for southern NEAC 1SW PFA and 13 for southern NEAC MSW PFA).

In 2013, the Working Group further assessed the effects of applying stricter criteria than $r^{2} \geq 0.2$ for inclusion in the FWI. As stricter criteria are used, the number of indicators included reduces rapidly (Figure 3.7.2.1). It was concluded to keep the criterion of $r^{2} \geq 0.2$ in order to obtain a sufficient number of indicators to be able to use the FWI even in the event of one or more indicators being unavailable by the time the FWI is applied each year. The $r^{2}$ value of 0.2 corresponds to a value slightly lower than what is considered to be a "large" effect size ( $r=0.5, r^{2}=0.25$ ) by Cohen (1988). Even though a criterion of $r^{2} \geq$ 0.2 gives each indicator little predictive power alone (Prairie, 1996), the approach of using a suite of indicators is more similar to metaanalysis (Rosenthal, 1984) meaning that the
outcome of the FWI is not dependent on the result of one indicator in isolation, but rather on the combined performance of the indicator set (see Annex 9).

The following summarizes the main steps performed by the spreadsheet following updating of the relevant data for the variable of interest by adding the latest year's number:

- Regression analysis with the dataset $x$ to determine its power to predict PFA in the forecasted years.
- Calculation of the $75 \%$ confidence intervals of individual predictions of the regression for dataset $x$. An indicator value below the $75 \%$ individual confidence interval ( CI ) is interpreted as indicative of an overestimation of the PFA, while a point above the $75 \%$ individual confidence interval is interpreted as indicative of an underestimation of PFA.
- A dataset is considered informative and should be kept as an indicator in the FWI if the following conditions are met: sample size $(n) \geq 10 ; r^{2} \geq 0.2$; dataset updated annually and new value available by January 15 . Datasets that do not meet these criteria are discarded.
- Apply a binary score to each indicator value. Thus, for dataset $x$, if the current year's indicator value is outside the $75 \%$ individual regression point estimate CI (below or above) then that indicator receives a score of 1 . If the indicator is within the $75 \% \mathrm{CI}$, then the indicator receives a score of -1 . In the absence of an indicator datapoint for any year, a score of zero is applied. Whether the indicator value is above or below the upper and lower CI values is checked separately in two spreadsheet columns and a decision whether the indicator value is within the CI is assessed by combining the information in the two columns.
- Separate columns are used to sum the scores for all the indicator datasets within each stock complex. This is done separately for points that fall above the CI and those that fall below. In the case of a two-sided approach (open fishery), if the sum of these columns is $\geq 0$, then the spreadsheet signals "REASSESS"; if the sum is $<0$, then it signals "No significant OVERestimation of PFA identified by indicators, do not reassess" for indicator values that fall below the CI, and "No significant UNDERestimation of PFA identified by indicators, do not reassess" for indicator values that are above the CI. In case of a one-sided approach (closed fishery), only underestimation will signal a "REASSESS".
- FWI results are generated for each stock complex (northern NEAC maturing and non-maturing, and southern NEAC maturing and non-maturing). A score of $\geq 0$ for any of these stock complexes would signal a reassessment.

Table 3.1.3.1 Number of gear units licensed or authorised by country and gear type.

| Year | $\begin{array}{cc} & \text { England \& } \\ \text { Gillnet } & \text { Sweepnet }\end{array}$ |  | Wales |  |  | UK (Scotland) |  | UK (N. Ireland) |  |  | Norway |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Gillnet <br> licences | Sweepnet | Hand-held net | Fixed engine | Rod \& Line | $\begin{aligned} & \hline \begin{array}{l} \text { Fixed } \\ \text { engine }^{1} \end{array} \end{aligned}$ | Net and coble ${ }^{2}$ | Driftnet | Draftnet | Bagnets and boxes | Bagnet | Bendnet | Liftnet | Driftnet (No. nets) |
| 1971 | 437 | 230 | 294 | 79 | - | 3080 | 800 | 142 | 305 | 18 | 4608 | 2421 | 26 | 8976 |
| 1972 | 308 | 224 | 315 | 76 | - | 3455 | 813 | 130 | 307 | 18 | 4215 | 2367 | 24 | 13448 |
| 1973 | 291 | 230 | 335 | 70 | - | 3256 | 891 | 130 | 303 | 20 | 4047 | 2996 | 32 | 18616 |
| 1974 | 280 | 240 | 329 | 69 | - | 3188 | 782 | 129 | 307 | 18 | 3382 | 3342 | 29 | 14078 |
| 1975 | 269 | 243 | 341 | 69 | - | 2985 | 773 | 127 | 314 | 20 | 3150 | 3549 | 25 | 15968 |
| 1976 | 275 | 247 | 355 | 70 | - | 2862 | 760 | 126 | 287 | 18 | 2569 | 3890 | 22 | 17794 |
| 1977 | 273 | 251 | 365 | 71 | - | 2754 | 684 | 126 | 293 | 19 | 2680 | 4047 | 26 | 30201 |
| 1978 | 249 | 244 | 376 | 70 | - | 2587 | 692 | 126 | 284 | 18 | 1980 | 3976 | 12 | 23301 |
| 1979 | 241 | 225 | 322 | 68 | - | 2708 | 754 | 126 | 274 | 20 | 1835 | 5001 | 17 | 23989 |
| 1980 | 233 | 238 | 339 | 69 | - | 2901 | 675 | 125 | 258 | 20 | 2118 | 4922 | 20 | 25652 |
| 1981 | 232 | 219 | 336 | 72 | - | 2803 | 655 | 123 | 239 | 19 | 2060 | 5546 | 19 | 24081 |
| 1982 | 232 | 221 | 319 | 72 | - | 2396 | 647 | 123 | 221 | 18 | 1843 | 5217 | 27 | 22520 |
| 1983 | 232 | 209 | 333 | 74 | - | 2523 | 668 | 120 | 207 | 17 | 1735 | 5428 | 21 | 21813 |
| 1984 | 226 | 223 | 354 | 74 | - | 2460 | 638 | 121 | 192 | 19 | 1697 | 5386 | 35 | 21210 |
| 1985 | 223 | 230 | 375 | 69 | - | 2010 | 529 | 122 | 168 | 19 | 1726 | 5848 | 34 | 20329 |
| 1986 | 220 | 221 | 368 | 64 | - | 1955 | 591 | 121 | 148 | 18 | 1630 | 5979 | 14 | 17945 |
| 1987 | 213 | 206 | 352 | 68 | - | 1679 | 564 | 120 | 119 | 18 | 1422 | 6060 | 13 | 17234 |
| 1988 | 210 | 212 | 284 | 70 | - | 1534 | 385 | 115 | 113 | 18 | 1322 | 5702 | 11 | 15532 |
| 1989 | 201 | 199 | 282 | 75 | - | 1233 | 353 | 117 | 108 | 19 | 1888 | 4100 | 16 | 0 |
| 1990 | 200 | 204 | 292 | 69 | - | 1282 | 340 | 114 | 106 | 17 | 2375 | 3890 | 7 | 0 |
| 1991 | 199 | 187 | 264 | 66 | - | 1137 | 295 | 118 | 102 | 18 | 2343 | 3628 | 8 | 0 |
| 1992 | 203 | 158 | 267 | 65 | - | 851 | 292 | 121 | 91 | 19 | 2268 | 3342 | 5 | 0 |
| 1993 | 187 | 151 | 259 | 55 | - | 903 | 264 | 120 | 73 | 18 | 2869 | 2783 | - | 0 |
| 1994 | 177 | 158 | 257 | 53 | 37278 | 749 | 246 | 119 | 68 | 18 | 2630 | 2825 |  | 0 |
| 1995 | 163 | 156 | 249 | 47 | 34941 | 729 | 222 | 122 | 68 | 16 | 2542 | 2715 | - | 0 |
| 1996 | 151 | 132 | 232 | 42 | 35281 | 643 | 201 | 117 | 66 | 12 | 2280 | 2860 |  | 0 |
| 1997 | 139 | 131 | 231 | 35 | 32781 | 680 | 194 | 116 | 63 | 12 | 2002 | 1075 |  | 0 |
| 1998 | 130 | 129 | 196 | 35 | 32525 | 542 | 151 | 117 | 70 | 12 | 1865 | 1027 |  | 0 |
| 1999 | 120 | 109 | 178 | 30 | 29132 | 406 | 132 | 113 | 52 | 11 | 1649 | 989 | - | 0 |
| 2000 | 110 | 103 | 158 | 32 | 30139 | 381 | 123 | 109 | 57 | 10 | 1557 | 982 |  | 0 |
| 2001 | 113 | 99 | 143 | 33 | 24350 | 387 | 95 | 107 | 50 | 6 | 1976 | 1081 |  | 0 |
| 2002 | 113 | 94 | 147 | 32 | 29407 | 426 | 102 | 106 | 47 | 4 | 1666 | 917 | - | 0 |
| 2003 | 58 | 96 | 160 | 57 | 29936 | 363 | 109 | 105 | 52 | 2 | 1664 | 766 |  | 0 |
| 2004 | 57 | 75 | 157 | 65 | 32766 | 450 | 118 | 90 | 54 | 2 | 1546 | 659 |  | 0 |
| 2005 | 59 | 73 | 148 | 65 | 34040 | 381 | 101 | 93 | 57 | 2 | 1453 | 661 |  | 0 |
| 2006 | 52 | 57 | 147 | 65 | 31606 | 364 | 86 | 107 | 49 | 2 | 1283 | 685 |  | 0 |
| 2007 | 53 | 45 | 157 | 66 | 32181 | 238 | 69 | 20 | 12 | 2 | 1302 | 669 |  | 0 |
| 2008 | 55 | 42 | 130 | 66 | 33900 | 181 | 77 | 20 | 12 | 2 | 957 | 653 |  | 0 |
| 2009 | 50 | 42 | 118 | 66 | 36461 | 162 | 64 | 20 | 12 | 2 | 978 | 631 |  | 0 |
| 2010 | 51 | 40 | 118 | 66 | 36159 | 189 | 66 | 2 | 1 | 2 | 760 | 493 |  | 0 |
| 2011 | 53 | 41 | 117 | 66 | 36991 | 201 | 74 | 2 | 1 | 2 | 767 | 506 | - | 0 |
| 2012 | 51 | 34 | 115 | 74 | 34665 | 251 | 83 | 1 | 1 |  | 749 | 448 | - | 0 |
| Mean 2007-2011 | 52 | 42 | 128 | 66 | 35060 | 194 | 70 | 13 | 8 | 2 | 953 | 590 |  | 0 |
| $\%$ change $^{3}$ | -2,7 | -19,0 | -10,2 | 12,1 | -1,1 | 29,3 | 19,4 | -92,2 | -86,8 | 0,0 | -21,4 | -24,1 |  |  |
| Mean 2002-2011 | 60 | 61 | 140 | 61 | 33345 | 295 | 86 | 57 | 30 | 2 | 1238 | 664 |  | 0 |
| $\%$ change $^{3}$ | -15,1 | $-43,8$ | -17,8 | 20,5 | 4,0 | -15,1 | -3,5 | -98,2 | -96,6 | -9,1 | -39,5 | -32,5 |  |  |

${ }^{1}$ Number of gear units expressed as trap months.
Number of gear units expressed as crew months.
${ }_{3}^{(2012 / \text { mean - 1) }}$ * 10
${ }^{4}$ Dash means "no data"

Table 3．1．3．1 Cont＇d．Number of gear units licensed or authorised by country and gear type．

| Year | Ireland |  |  |  | Finland |  |  |  | France |  |  | Russia |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Driftnets No． |  | Other nets Commercial | Rod | The Teno River |  |  | $\begin{aligned} & \hline \begin{array}{l} \text { R. Näätämön } \\ \hline \text { Recreational } \\ \text { fishery } \\ \hline \end{array} ⿳ ⺈ ⿴ 囗 十 一 \text {. } \end{aligned}$ | Rod and line licences in freshwater | Com．nets in freshwater ${ }^{\text {1a }}$ | Drift net <br> Licences in estuary ${ }^{1 \mathrm{~b},{ }^{2}}$ | Kola Peninsula <br> Catch－and－release C <br> Fishing days | Archangel region Commercial， number of gears Coastal |  |
|  |  | Draftnets |  |  | Recreational fishery <br> Tourist anglers |  | Local rod and net fishery |  |  |  |  |  |  |  |
|  |  |  |  |  | Fishing days | Fishermen | Fishermen | Fishermen |  |  |  |  |  | In－river |
| 1971 | 916 | 697 | 213 | 10566 |  | － | － |  | － | － |  |  | － |  |
| 1972 | 1156 | 678 | 197 | 9612 | － |  | － | － | － | － | － |  | － |  |
| 1973 | 1112 | 713 | 224 | 11660 | － | － | － | － | － | － | － | － | － |  |
| 1974 | 1048 | 681 | 211 | 12845 | － | － | － | － | － | － | － | － | － |  |
| 1975 | 1046 | 672 | 212 | 13142 | － | － | － | － | － | － | － | － | － |  |
| 1976 | 1047 | 677 | 225 | 14139 | － | － | － | － | － | － | － | － | － |  |
| 1977 | 997 | 650 | 211 | 11721 | － | － | － | － | － | － | － | － | － |  |
| 1978 | 1007 | 608 | 209 | 13327 | － | － | － | － | － | － | － | － | － |  |
| 1979 | 924 | 657 | 240 | 12726 | － | － | － | － | － | － | － | － | － | － |
| 1980 | 959 | 601 | 195 | 15864 | － | － | － | － | － | － | － | － | － |  |
| 1981 | 878 | 601 | 195 | 15519 | 16859 | 5742 | 677 | 467 | － | － | － | － | － |  |
| 1982 | 830 | 560 | 192 | 15697 | 19690 | 7002 | 693 | 484 | 4145 | 55 | 82 | － | － | － |
| 1983 | 801 | 526 | 190 | 16737 | 20363 | 7053 | 740 | 587 | 3856 | 49 | 82 | － | － |  |
| 1984 | 819 | 515 | 194 | 14878 | 21149 | 7665 | 737 | 677 | 3911 | 42 | 82 | － | － |  |
| 1985 | 827 | 526 | 190 | 15929 | 21742 | 7575 | 740 | 866 | 4443 | 40 | 82 | － | － | － |
| 1986 | 768 | 507 | 183 | 17977 | 21482 | 7404 | 702 | 691 | 5919 | $58^{3}$ | 86 | － | － | － |
| 1987 | 768 | 507 | 183 | 17977 | 22487 | 7759 | 754 | 689 | $5724{ }^{4}$ | $87^{4}$ | 80 | － | － |  |
| 1988 | 836 | 507 | 183 | 11539 | 21708 | 7755 | 741 | 538 | 4346 | 101 | 76 | － | － |  |
| 1989 | 801 | 507 | 183 | 16484 | 24118 | 8681 | 742 | 696 | 3789 | 83 | 78 | － | － |  |
| 1990 | 756 | 525 | 189 | 15395 | 19596 | 7677 | 728 | 614 | 2944 | 71 | 76 | － | － |  |
| 1991 | 707 | 504 | 182 | 15178 | 22922 | 8286 | 734 | 718 | 2737 | 78 | 71 | 1711 | － |  |
| 1992 | 691 | 535 | 183 | 20263 | 26748 | 9058 | 749 | 875 | 2136 | 57 | 71 | 4088 | － | － |
| 1993 | 673 | 457 | 161 | 23875 | 29461 | 10198 | 755 | 705 | 2104 | 53 | 55 | 6026 | 59 | 199 |
| 1994 | 732 | 494 | 176 | 24988 | 26517 | 8985 | 751 | 671 | 1672 | 14 | 59 | 8619 | 60 | 230 |
| 1995 | 768 | 512 | 164 | 27056 | 24951 | 8141 | 687 | 716 | 1878 | 17 | 59 | 5822 | 55 | 239 |
| 1996 | 778 | 523 | 170 | 29759 | 17625 | 5743 | 672 | 814 | 1798 | 21 | 69 | 6326 | 85 | 330 |
| 1997 | 852 | 531 | 172 | 31873 | 16255 | 5036 | 616 | 588 | 2953 | 10 | 59 | 6355 | 68 | 282 |
| 1998 | 874 | 513 | 174 | 31565 | 18700 | 5759 | 621 | 673 | 2352 | 16 | 63 | 6034 | 66 | 270 |
| 1999 | 874 | 499 | 162 | 32493 | 22935 | 6857 | 616 | 850 | 2225 | 15 | 61 | 7023 | 66 | 194 |
| 2000 | 871 | 490 | 158 | 33527 | 28385 | 8275 | 633 | 624 | $2037{ }^{5}$ | 16 | 51 | 7336 | 60 | 173 |
| 2001 | 881 | 540 | 155 | 32814 | 33501 | 9367 | 863 | 590 | 2080 | 18 | 63 | 8468 | 53 | 121 |
| 2002 | 833 | 544 | 159 | 35024 | 37491 | 10560 | 853 | 660 | 2082 | 18 | 65 | 9624 | 63 | 72 |
| 2003 | 877 | 549 | 159 | 31809 | 34979 | 10032 | 832 | 644 | 2048 | 18 | 60 | 11994 | 55 | 84 |
| 2004 | 831 | 473 | 136 | 30807 | 29494 | 8771 | 801 | 657 | 2158 | 15 | 62 | 13300 | 62 | 56 |
| 2005 | 877 | 518 | 158 | 28738 | 27627 | 7776 | 785 | 705 | 2356 | 16 | 59 | 20309 | 93 | 69 |
| 2006 | 875 | 533 | 162 | 27341 | 29516 | 7749 | 836 | 552 | 2269 | 12 | 57 | 13604 | 62 | 72 |
| 2007 | 0 | 335 | 100 | 19986 | 33664 | 8763 | 780 | 716 | 2431 | 13 | 59 | n／a | 82 | 53 |
| 2008 | 0 | 160 | 0 | 20061 | 31143 | 8111 | 756 | 694 | 2401 | 12 | 56 | n／a | 66 | 62 |
| 2009 | 0 | 146 | 38 | 18314 | 29641 | 7676 | 761 | 656 | 2421 | 12 | 37 | n／a | 79 | 72 |
| 2010 | 0 | 166 | 40 | 17983 | 30646 | 7814 | 756 | 615 | 2200 | 12 | 33 | n／a | 55 | 66 |
| 2011 | 0 | 154 | 91 | 19899 | 31269 | 7915 | 776 | 767 | 2540 | 12 | 29 | n／a | 78 | 52 |
| 2012 |  | 149 | 86 | 19588 | 32614 | 7930 | 785 | 781 | 2799 | 12 | 25 | n／a | 72 | 53 |
| Mean 2007－2011 | 0 | 192 | 54 | 19249 | 31273 | 8056 | 766 | 690 | 2399 | 12 | 43 |  | 72 | 61 |
| $\%$ change ${ }^{6}$ | 0，0 | －19，9 | 69，1 | 3，4 | 0，0 | －1，7 | 1，3 | 11，2 | 5，9 | －1，6 | －32，2 |  | 8,3 | －14，8 |
| Mean 2002－2011 | 429 | 358 | 104 | 24996 | 31547 | 8517 | 794 | 667 | 2291 | 14 | 52 | 13766 | 70 | 66 |
| $\%$ change ${ }^{6}$ | －100， 0 | －57，0 | －12，8 | －20，4 | －0，9 | －7，1 | $-2,2$ | 15，1 | 10，9 | －14，3 | －43，9 |  | 12，2 | $-21,0$ |

[^0]Table 3.1.4.1 Nominal catch of SALMON in NEAC Area (in tonnes round fresh weight), 1960-2012 (2012 figures are provisional).

| Year | Southern countries | Northern countries <br> (1) | Faroes(2) | Other catches in international waters | Total Reported Catch | Unreported catches |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | $\begin{gathered} \hline \text { NEAC } \\ \text { Area (3) } \\ \hline \end{gathered}$ | International waters (4) |
| 1960 | 2641 | 2899 | - | - | 5540 | - | - |
| 1961 | 2276 | 2477 | - | - | 4753 | - | - |
| 1962 | 3894 | 2815 | - | - | 6709 | - | - |
| 1963 | 3842 | 2434 | - | - | 6276 | - | - |
| 1964 | 4242 | 2908 | - | - | 7150 | - | - |
| 1965 | 3693 | 2763 | - | - | 6456 | - | - |
| 1966 | 3549 | 2503 | - | - | 6052 | - | - |
| 1967 | 4492 | 3034 | - | - | 7526 | - | - |
| 1968 | 3623 | 2523 | 5 | 403 | 6554 | - | - |
| 1969 | 4383 | 1898 | 7 | 893 | 7181 | - | - |
| 1970 | 4048 | 1834 | 12 | 922 | 6816 | - | - |
| 1971 | 3736 | 1846 | - | 471 | 6053 | - | - |
| 1972 | 4257 | 2340 | 9 | 486 | 7092 | - | - |
| 1973 | 4604 | 2727 | 28 | 533 | 7892 | - | - |
| 1974 | 4352 | 2675 | 20 | 373 | 7420 | - | - |
| 1975 | 4500 | 2616 | 28 | 475 | 7619 | - | - |
| 1976 | 2931 | 2383 | 40 | 289 | 5643 | - | - |
| 1977 | 3025 | 2184 | 40 | 192 | 5441 | - | - |
| 1978 | 3102 | 1864 | 37 | 138 | 5141 | - | - |
| 1979 | 2572 | 2549 | 119 | 193 | 5433 | - | - |
| 1980 | 2640 | 2794 | 536 | 277 | 6247 | - | - |
| 1981 | 2557 | 2352 | 1025 | 313 | 6247 | - | - |
| 1982 | 2533 | 1938 | 606 | 437 | 5514 | - | - |
| 1983 | 3532 | 2341 | 678 | 466 | 7017 | - | - |
| 1984 | 2308 | 2461 | 628 | 101 | 5498 | - | - |
| 1985 | 3002 | 2531 | 566 | - | 6099 | - | - |
| 1986 | 3595 | 2588 | 530 | - | 6713 | - | - |
| 1987 | 2564 | 2266 | 576 | - | 5406 | 2554 | - |
| 1988 | 3315 | 1969 | 243 | - | 5527 | 3087 | - |
| 1989 | 2433 | 1627 | 364 | - | 4424 | 2103 | - |
| 1990 | 1645 | 1775 | 315 | - | 3735 | 1779 | 180-350 |
| 1991 | 1145 | 1677 | 95 | - | 2917 | 1555 | 25-100 |
| 1992 | 1523 | 1806 | 23 | - | 3352 | 1825 | 25-100 |
| 1993 | 1443 | 1853 | 23 | - | 3319 | 1471 | 25-100 |
| 1994 | 1896 | 1684 | 6 | - | 3586 | 1157 | 25-100 |
| 1995 | 1775 | 1503 | 5 | - | 3283 | 942 | - |
| 1996 | 1392 | 1358 | - | - | 2750 | 947 | - |
| 1997 | 1112 | 962 | - | - | 2074 | 732 | - |
| 1998 | 1120 | 1099 | 6 | - | 2225 | 1108 | - |
| 1999 | 934 | 1139 | 0 | - | 2073 | 887 | - |
| 2000 | 1210 | 1518 | 8 | - | 2736 | 1135 | - |
| 2001 | 1242 | 1634 | 0 | - | 2876 | 1089 | - |
| 2002 | 1135 | 1360 | 0 | - | 2495 | 946 | - |
| 2003 | 908 | 1394 | 0 | - | 2302 | 719 | - |
| 2004 | 919 | 1058 | 0 | - | 1977 | 575 | - |
| 2005 | 809 | 1189 | 0 | - | 1998 | 605 | - |
| 2006 | 650 | 1217 | 0 | - | 1867 | 604 | - |
| 2007 | 373 | 1036 | 0 | - | 1409 | 465 | - |
| 2008 | 355 | 1178 | 0 | - | 1533 | 433 | - |
| 2009 | 265 | 898 | 0 | - | 1163 | 317 | - |
| 2010 | 411 | 1003 | 0 | - | 1415 | 357 | - |
| 2011 | 410 | 1009 | 0 | - | 1419 | 382 |  |
| 2012 | 301 | 939 | 0 | - | 1240 | 363 |  |
| Average |  |  |  |  |  |  |  |
| 2007-2011 | 363 | 1025 | 0 | - | 1388 | 391 | - |
| 2002-2011 | 624 | 1134 | 0 | - | 1758 | 540 | - |

1. All Iceland has been included in Northern countries
2. Since 1991, fishing carried out at the Faroes has only been for research purposes.
3. No unreported catch estimate available for Russia since 2008.
4. Estimates refer to season ending in given year.

Table 3.1.5.1 CPUE for salmon rod fisheries in Finland (Teno, Naatamo), France, and UK(N.Ireland)(Bush).

|  | Finland (R. Teno) |  | Finland (R. Naatamo) |  | France | UK(N.Ire.)(R.Bush) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Catch per | Catch per | Catch per | Catch per | Catch per | Catch per |  |
|  | angler seaso | angler day | angler season | angler day | angler season | rod day |  |
| Year | kg | kg | kg | kg | Number | Number |  |
|  |  |  |  |  |  |  |  |
| 1974 |  | 2.8 |  |  |  |  |  |
| 1975 |  | 2.7 |  |  |  |  |  |
| 1976 |  | - |  |  |  |  |  |
| 1977 |  | 1.4 |  |  |  |  |  |
| 1978 |  | 1.1 |  |  |  |  |  |
| 1979 |  | 0.9 |  |  |  |  |  |
| 1980 |  | 1.1 |  |  |  |  |  |
| 1981 | 3.2 | 1.2 |  |  |  |  |  |
| 1982 | 3.4 | 1.1 |  |  |  |  |  |
| 1983 | 3.4 | 1.2 |  |  |  | 0.248 |  |
| 1984 | 2.2 | 0.8 | 0.5 | 0.2 |  | 0.083 |  |
| 1985 | 2.7 | 0.9 | n/a | n/a |  | 0.283 |  |
| 1986 | 2.1 | 0.7 | n/a | n/a |  | 0.274 |  |
| 1987 | 2.3 | 0.8 | n/a | n/a | 0.39 | 0.194 |  |
| 1988 | 1.9 | 0.7 | 0.5 | 0.2 | 0.73 | 0.165 |  |
| 1989 | 2.2 | 0.8 | 1.0 | 0.4 | 0.55 | 0.135 |  |
| 1990 | 2.8 | 1.1 | 0.7 | 0.3 | 0.71 | 0.247 |  |
| 1991 | 3.4 | 1.2 | 1.3 | 0.5 | 0.60 | 0.396 |  |
| 1992 | 4.5 | 1.5 | 1.4 | 0.3 | 0.94 | 0.258 |  |
| 1993 | 3.9 | 1.3 | 0.4 | 0.2 | 0.88 | 0.341 |  |
| 1994 | 2.4 | 0.8 | 0.6 | 0.2 | 2.32 | 0.205 |  |
| 1995 | 2.7 | 0.9 | 0.5 | 0.1 | 1.15 | 0.206 |  |
| 1996 | 3.0 | 1.0 | 0.7 | 0.2 | 1.57 | 0.267 |  |
| 1997 | 3.4 | 1.0 | 1.1 | 0.2 | 0.44 | 0.338 |  |
| 1998 | 3.0 | 0.9 | 1.3 | 0.3 | 0.67 | 0.569 |  |
| 1999 | 3.7 | 1.1 | 0.8 | 0.2 | 0.76 | 0.273 |  |
| 2000 | 5.0 | 1.5 | 0.9 | 0.2 | 1.06 | 0.259 |  |
| 2001 | 5.9 | 1.7 | 1.2 | 0.3 | 0.97 | 0.444 |  |
| 2002 | 3.1 | 0.9 | 0.7 | 0.2 | 0.84 | 0.184 |  |
| 2003 | 2.6 | 0.7 | 0.8 | 0.2 | 0.76 | 0.238 |  |
| 2004 | 1.4 | 0.4 | 0.9 | 0.2 | 1.25 | 0.252 |  |
| 2005 | 2.7 | 0.8 | 1.3 | 0.2 | 0.74 | 0.323 |  |
| 2006 | 3.4 | 1.0 | 1.9 | 0.4 | 0.89 | 0.457 |  |
| 2007 | 2.9 | 0.8 | 1.0 | 0.2 | 0.74 | 0.601 |  |
| 2008 | 4.2 | 1.1 | 0.9 | 0.2 | 0.77 | 0.457 |  |
| 2009 | 2.3 | 0.6 | 0.7 | 0.1 | 0.50 | 0.136 |  |
| 2010 | 3.0 | 0.8 | 1.3 | 0.2 | 0.87 | 0.226 |  |
| 2011 | 2.4 | 0.6 | 1.0 | 0.2 | 0.65 | 0.122 |  |
| 2012 | 3.6 | 0.9 | 1.7 | 0.4 | 0.61 | 0.149 |  |
| Mean |  |  |  |  |  |  |  |
| 2007-11 | 3.0 | 0.8 | 1.0 | 0.2 | 0.7 | 0.3 |  |
|  |  |  |  |  |  |  |  |

${ }^{1}$ Large numbers of new, inexperienced anglers in 1997 because cheaper licence types were introduced.

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Table 3.1.5.2 CPUE for salmon rod fisheries in the Barents Sea and White Sea basin and commercial catches in Archangelsk region in Russia.

|  |  |  |  |  |  |  |  |  |  | Archangelsk region <br> Commercial fishery (tonnes/gear, |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Barents Sea Basin, catch per angler day |  |  |  | Iokanga | White Sea Basin, catch per angler day |  |  |  | Commercial fishery (tonnes/gear |  |
| Year | Rynda | Kharlovka | E. Litsa | Varzina |  | Ponoi | Varzuga | Kitsa | Umba | Coastal | In-river |
|  |  |  |  |  |  |  |  |  |  |  |  |
| 1991 |  |  |  |  |  | 2.79 | 1.87 |  | 1.33 |  |  |
| 1992 | 2.37 | 1.45 | 2.95 | 1.07 | 0.14 | 4.50 | 2.26 | 1.21 | 1.37 |  |  |
| 1993 | 1.18 | 1.46 | 1.59 | 0.49 | 0.65 | 3.57 | 1.28 | 1.43 | 2.72 | 0.34 | 0.04 |
| 1994 | 0.71 | 0.85 | 0.79 | 0.55 | 0.33 | 3.30 | 1.60 | 1.59 | 1.44 | 0.35 | 0.05 |
| 1995 | 0.49 | 0.78 | 0.94 | 1.22 | 0.72 | 3.77 | 2.52 | 1.78 | 1.20 | 0.22 | 0.08 |
| 1996 | 0.70 | 0.85 | 1.31 | 1.50 | 1.40 | 3.78 | 1.44 | 1.76 | 0.93 | 0.19 | 0.02 |
| 1997 | 1.20 | 0.71 | 1.09 | 0.61 | 1.41 | 6.09 | 2.36 | 2.48 | 1.46 | 0.23 | 0.02 |
| 1998 | 1.01 | 0.55 | 0.75 | 0.44 | 0.87 | 4.52 | 2.28 | 2.78 | 0.98 | 0.24 | 0.03 |
| 1999 | 0.95 | 0.77 | 0.93 | 0.43 | 1.19 | 3.30 | 1.71 | 1.66 | 0.76 | 0.22 | 0.04 |
| 2000 | 1.35 | 0.77 | 0.89 | 0.57 | 2.28 | 3.55 | 1.53 | 3.02 | 1.25 | 0.28 | 0.03 |
| 2001 | 1.48 | 0.92 | 1.00 | 0.89 | 0.73 | 4.35 | 1.86 | 1.81 | 1.04 | 0.21 | 0.04 |
| 2002 | 2.39 | 0.99 | 0.89 | 0.80 | 2.82 | 7.28 | 1.44 | 2.11 | 0.36 | 0.21 | 0.11 |
| 2003 | 1.61 | 1.14 | 1.04 | 0.79 | 2.01 | 8.39 | 1.17 | 1.61 | 0.36 | 0.16 | 0.05 |
| 2004 | 1.07 | 0.98 | 1.31 | 0.65 | 1.00 | 5.80 | 1.14 | 1.10 | 0.36 | 0.25 | 0.08 |
| 2005 | 1.09 | 0.82 | 1.45 | 0.46 | 0.88 | 4.42 | 0.57 | 0.89 | 0.28 | 0.17 | 0.08 |
| 2006 | 0.98 | 1.49 | 1.49 | 1.45 |  | 6.28 | 2.23 |  | 0.73 | 0.19 | 0.05 |
| 2007 | 0.92 | 0.78 | 1.43 | 1.16 |  | 5.96 |  |  |  | 0.14 | 0.09 |
| 2008 |  |  |  |  |  | 5.73 |  |  |  | 0.12 | 0.08 |
| 2009 |  |  |  |  |  | 5.72 |  |  |  | 0.09 | 0.05 |
| 2010 |  |  |  |  |  | 4.78 |  |  |  | 0.21 | 0.08 |
| 2011 |  |  |  |  |  | 4.01 |  |  |  | 0.15 | 0.07 |
| 2012 |  |  |  |  |  | 5.56 |  |  |  | 0.17 | 0.09 |
| Mean |  |  |  |  |  |  |  |  |  |  |  |
| Data available | 1.22 | 0.96 | 1.24 | 0.82 | 1.17 | 4.89 | 1.70 | 1.80 | 1.03 | 0.21 | 0.06 |
| 2007-11 |  |  |  |  |  | 5.24 |  |  |  | 0.14 | 0.07 |


| Table 3.1.5.3 |  | CPUE data for net and fixed engine salmon fisheries by Region in UK |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | (England \& Wales). Data expressed as catch per licence-tide, |  |  |  |  |  |
|  |  | except NE, for which the data are recorded as catch per licence-day. |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
|  |  | Region (aggregated data, various methods) |  |  |  |  |  |
|  | North East |  |  |  |  |  |  |
| Year | drift nets | North East | South West | Midlands | Wales | North West |  |
| 1988 |  | 5.49 |  |  |  | - |  |
| 1989 |  | 4.39 |  |  |  | 0.82 |  |
| 1990 |  | 5.53 |  |  |  | 0.63 |  |
| 1991 |  | 3.20 |  |  |  | 0.51 |  |
| 1992 |  | 3.83 |  |  |  | 0.40 |  |
| 1993 | 8.23 | 6.43 |  |  |  | 0.63 |  |
| 1994 | 9.02 | 7.53 |  |  |  | 0.71 |  |
| 1995 | 11.18 | 7.84 |  |  |  | 0.79 |  |
| 1996 | 4.93 | 3.74 |  |  |  | 0.59 |  |
| 1997 | 6.48 | 4.40 | 0.70 | 0.48 | 0.07 | 0.63 |  |
| 1998 | 5.92 | 3.81 | 1.25 | 0.42 | 0.08 | 0.46 |  |
| 1999 | 8.06 | 4.88 | 0.79 | 0.72 | 0.02 | 0.52 |  |
| 2000 | 13.06 | 8.11 | 1.01 | 0.66 | 0.18 | 1.05 |  |
| 2001 | 10.34 | 6.83 | 0.71 | 0.79 | 0.16 | 0.71 |  |
| 2002 | 8.55 | 5.59 | 1.03 | 1.39 | 0.23 | 0.90 |  |
| 2003 | 7.13 | 4.82 | 1.24 | 1.13 | 0.11 | 0.62 |  |
| 2004 | 8.17 | 5.88 | 1.17 | 0.46 | 0.11 | 0.69 |  |
| 2005 | 7.23 | 4.13 | 0.60 | 0.97 | 0.09 | 1.28 |  |
| 2006 | 5.60 | 3.20 | 0.66 | 0.97 | 0.09 | 0.82 |  |
| 2007 | 7.24 | 4.17 | 0.33 | 1.26 | 0.05 | 0.75 |  |
| 2008 | 5.41 | 3.59 | 0.63 | 1.33 | 0.06 | 0.34 |  |
| 2009 | 4.76 | 3.08 | 0.53 | 1.67 | 0.04 | 0.51 |  |
| 2010 | 17.03 | 8.56 | 0.99 | 0.26 | 0.09 | 0.47 |  |
| 2011 | 19.25 | 9.93 | 0.63 | 0.14 | 0.10 | 0.34 |  |
| 2012 | 7.01 | 5.53 | 0.69 | n/a | 0.21 | 0.31 |  |
| Mean |  |  |  |  |  |  |  |
| 2007-11 | 10.74 | 5.87 | 0.62 | 0.93 | 0.07 | 0.48 |  |

Table 3.1.5.4 Catch per unit of effort (CPUE) for salmon rod fisheries in each Region in UK (England \& Wales), 1997-2012.
[CPUE is expressed as number of salmon (including released fish) caught per 100 days fished].


Table 3.1.5.5 CPUE data for Scottish net fisheries.
Catch in numbers of fish per unit effort.

| Year | Fixed engine |
| :---: | :---: |
|  | Catch/trap month ${ }^{1}$ |
|  | Catch/crew month |


| 1 |
| :--- |
| 1 |
| 1 |
| 1 |
| 1 |
| 1 |
| 1 |
| 1 |
| 1 |
| 1 |

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Table 3.1.5.6 Catch per unit effort for the marine fishery in Norway. The CPUE is expressed as numbers of salmon caught per net day in bagnets and bendnets

|  | divided by salmon weight. |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Bagnet |  |  | Bendnet |  |
| Year | $<3 \mathrm{~kg}$ | $3-7 \mathrm{~kg}$ | $>7 \mathrm{~kg}$ | $<3 \mathrm{~kg}$ | 3-7 kg | $>7 \mathrm{~kg}$ |
| 1998 | 0.88 | 0.66 | 0.12 | 0.80 | 0.56 | 0.13 |
| 1999 | 1.16 | 0.72 | 0.16 | 0.75 | 0.67 | 0.17 |
| 2000 | 2.01 | 0.90 | 0.17 | 1.24 | 0.87 | 0.17 |
| 2001 | 1.52 | 1.03 | 0.22 | 1.03 | 1.39 | 0.36 |
| 2002 | 0.91 | 1.03 | 0.26 | 0.74 | 0.87 | 0.32 |
| 2003 | 1.57 | 0.90 | 0.26 | 0.84 | 0.69 | 0.28 |
| 2004 | 0.89 | 0.97 | 0.25 | 0.59 | 0.60 | 0.17 |
| 2005 | 1.17 | 0.81 | 0.27 | 0.72 | 0.73 | 0.33 |
| 2006 | 1.02 | 1.33 | 0.27 | 0.72 | 0.86 | 0.29 |
| 2007 | 0.43 | 0.90 | 0.32 | 0.57 | 0.95 | 0.33 |
| 2008 | 1.07 | 1.13 | 0.43 | 0.57 | 0.97 | 0.57 |
| 2009 | 0.73 | 0.92 | 0.31 | 0.44 | 0.78 | 0.32 |
| 2010 | 1.46 | 1.13 | 0.39 | 0.82 | 1.00 | 0.38 |
| 2011 | 1.30 | 1.98 | 0.35 | 0.71 | 1.02 | 0.36 |
| 2012 | 1.12 | 1.26 | 0.43 | 0.89 | 1.03 | 0.41 |
| Mean |  |  |  |  |  |  |
| 2007-11 | 1.00 | 1.21 | 0.36 | 0.62 | 0.94 | 0.39 |

Table 3.1.6.1. Percentage of 1SW salmon in catches from countries in the North East Atlantic, 1987-2012.

| Year | Iceland | Finland | Norway | Russia | Sweden | Northern countries | UK (Scot) | UK (E\&W) | France <br> (1) | Spain <br> (2) | Southern countries |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 |  | 66 | 61 | 71 |  | 63 | 61 | 68 | 77 |  | 63 |
| 1988 |  | 63 | 64 | 53 |  | 62 | 57 | 69 | 29 |  | 60 |
| 1989 | 69 | 66 | 73 | 73 | 41 | 72 | 63 | 65 | 33 |  | 63 |
| 1990 | 66 | 64 | 68 | 73 | 70 | 69 | 48 | 52 | 45 |  | 49 |
| 1991 | 71 | 59 | 65 | 70 | 71 | 66 | 53 | 71 | 39 |  | 58 |
| 1992 | 72 | 70 | 62 | 72 | 68 | 65 | 55 | 77 | 48 |  | 59 |
| 1993 | 76 | 58 | 61 | 61 | 62 | 63 | 57 | 81 | 74 | 64 | 64 |
| 1994 | 63 | 55 | 68 | 69 | 64 | 67 | 54 | 77 | 55 | 69 | 61 |
| 1995 | 71 | 59 | 58 | 70 | 78 | 62 | 53 | 72 | 60 | 26 | 59 |
| 1996 | 73 | 79 | 53 | 80 | 63 | 61 | 53 | 65 | 51 | 34 | 56 |
| 1997 | 73 | 69 | 64 | 82 | 54 | 68 | 54 | 73 | 51 | 28 | 60 |
| 1998 | 82 | 75 | 66 | 82 | 59 | 70 | 58 | 82 | 71 | 54 | 65 |
| 1999 | 70 | 83 | 65 | 78 | 71 | 68 | 45 | 68 | 27 | 14 | 54 |
| 2000 | 82 | 71 | 67 | 75 | 69 | 69 | 54 | 79 | 58 | 74 | 65 |
| 2001 | 78 | 48 | 58 | 74 | 55 | 60 | 55 | 75 | 51 | 40 | 62 |
| 2002 | 83 | 34 | 49 | 70 | 63 | 54 | 54 | 76 | 69 | 38 | 64 |
| 2003 | 75 | 51 | 61 | 67 | 47 | 62 | 52 | 66 | 51 | 16 | 55 |
| 2004 | 86 | 47 | 52 | 68 | 52 | 58 | 51 | 81 | 40 | 67 | 59 |
| 2005 | 87 | 72 | 67 | 66 | 55 | 69 | 58 | 76 | 41 | 15 | 61 |
| 2006 | 84 | 73 | 54 | 77 | 56 | 60 | 57 | 78 | 50 | 15 | 61 |
| 2007 | 91 | 30 | 42 | 69 | 33 | 50 | 57 | 78 | 45 | 26 | 61 |
| 2008 | 90 | 34 | 46 | 58 | 30 | 54 | 48 | 76 | 42 | 11 | 55 |
| 2009 | 91 | 62 | 49 | 63 | 34 | 59 | 49 | 72 | 42 | 30 | 54 |
| 2010 | 82 | 50 | 56 | 58 | 41 | 60 | 55 | 78 | 67 | 32 | 63 |
| 2011 | 85 | 61 | 41 | 58 | 32 | 49 | 36 | 57 | 35 | 2 | 45 |
| 2012 | 87 | 73 | 47 | 70 | 28 | 55 | 50 | 51 | 38 | 18 | 49 |
| Means |  |  |  |  |  |  |  |  |  |  |  |
| 2007-2011 | 88 | 47 | 47 | 61 | 34 | 55 | 49 | 72 | 46 | 20 | 56 |
| 2002-2011 | 85 | 52 | 52 | 65 | 44 | 58 | 52 | 74 | 48 | 25 | 58 |



Table 3.3.5.1
Estimated number of RETURNING 1SW salmon by NEAC country or region and year

| Year | Northern Europe |  |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  |  |  | NEAC Area |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finland | Iceland | Norway | Russia | Sweden | Total |  |  | France | $\begin{array}{\|c\|} \hline \text { Iceland } \\ \hline \text { S\&W } \end{array}$ | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  |  | Total |  |  |
|  |  | N\&E |  |  |  | 5.0\% | 50.0\% | 95.0\% |  |  |  |  |  |  | 5.0\% | 50.0\% | 95.0\% | 5.0\% | 50.0\% | 95.0\% |
| 1971 | 25,922 | 9,401 |  | 154,435 | 17,451 |  |  |  | 50,011 | 62,520 | 1,055,511 | 82,561 | 181,588 | 620,892 | 1,835,337 | 2,064,901 | 2,354,934 |  |  |  |
| 1972 | 40,474 | 8,620 |  | 117,498 | 13,816 |  |  |  | 100,300 | 50,769 | 1,124,390 | 79,581 | 158,788 | 542,701 | 1,832,389 | 2,073,072 | 2,379,487 |  |  |  |
| 1973 | 36,851 | 10,338 |  | 173,692 | 17,158 |  |  |  | 60,473 | 54,316 | 1,222,672 | 93,855 | 138,706 | 651,601 | 1,970,596 | 2,238,654 | 2,572,489 |  |  |  |
| 1974 | 73,075 | 10,269 |  | 172,536 | 24,646 |  |  |  | 28,248 | 38,774 | 1,391,967 | 117,252 | 151,559 | 620,088 | 2,072,457 | 2,359,710 | 2,727,432 |  |  |  |
| 1975 | 50,868 | 12,567 |  | 264,634 | 26,587 |  |  |  | 56,590 | 59,911 | 1,540,374 | 120,126 | 124,462 | 505,054 | 2,107,267 | 2,419,250 | 2,824,148 |  |  |  |
| 1976 | 34,955 | 12,588 |  | 183,990 | 14,986 |  |  |  | 51,724 | 47,345 | 1,047,950 | 80,333 | 86,710 | 433,790 | 1,538,875 | 1,758,352 | 2,040,583 |  |  |  |
| 1977 | 17,910 | 17,539 |  | 117,481 | 7,092 |  |  |  | 40,069 | 48,547 | 903,324 | 91,313 | 85,298 | 453,323 | 1,440,088 | 1,630,121 | 1,876,174 |  |  |  |
| 1978 | 24,399 | 17,803 |  | 118,893 | 8,079 |  |  |  | 41,004 | 63,630 | 789,930 | 104,459 | 111,265 | 519,911 | 1,466,326 | 1,640,214 | 1,858,218 |  |  |  |
| 1979 | 28,512 | 17,085 |  | 165,134 | 8,521 |  |  |  | 47,289 | 58,896 | 730,271 | 99,469 | 77,834 | 428,293 | 1,292,097 | 1,452,784 | 1,650,180 |  |  |  |
| 1980 | 12,804 | 2,583 |  | 117,131 | 10,763 |  |  |  | 97,876 | 26,671 | 553,289 | 93,387 | 98,889 | 266,489 | 1,025,807 | 1,149,538 | 1,305,021 |  |  |  |
| 1981 | 19,720 | 13,309 |  | 96,542 | 19,571 |  |  |  | 76,975 | 34,540 | 290,830 | 98,209 | 77,376 | 328,725 | 839,646 | 916,801 | 1,004,447 |  |  |  |
| 1982 | 5,790 | 6,145 |  | 84,867 | 17,225 |  |  |  | 48,380 | 35,380 | 604,459 | 83,197 | 111,923 | 472,717 | 1,240,741 | 1,364,626 | 1,506,568 |  |  |  |
| 1983 | 28,560 | 9,089 | 699,564 | 141,907 | 22,879 | 809,162 | 904,551 | 1,013,270 | 51,445 | 44,752 | 1,066,455 | 121,959 | 156,858 | 482,392 | 1,742,051 | 1,935,984 | 2,163,466 | 2,619,032 | 2,840,069 | 3,093,756 |
| 1984 | 31,851 | 3,288 | 728,594 | 152,732 | 32,202 | 850,252 | 951,397 | 1,069,869 | 84,664 | 27,538 | 559,007 | 106,650 | 61,670 | 510,099 | 1,237,135 | 1,361,569 | 1,502,930 | 2,154,026 | 2,317,161 | 2,492,987 |
| 1985 | 48,122 | 22,745 | 741,727 | 209,048 | 38,476 | 962,716 | 1,065,049 | 1,176,868 | 31,688 | 44,647 | 928,636 | 106,994 | 80,053 | 422,523 | 1,453,751 | 1,620,802 | 1,827,843 | 2,491,693 | 2,690,540 | 2,920,016 |
| 1986 | 43,822 | 28,263 | 646,466 | 179,363 | 40,496 | 854,523 | 941,727 | 1,040,146 | 48,380 | 73,341 | 1,039,344 | 123,926 | 89,812 | 523,701 | 1,717,952 | 1,915,227 | 2,151,951 | 2,644,946 | 2,859,648 | 3,113,772 |
| 1987 | 55,804 | 16,590 | 543,391 | 191,351 | 32,774 | 768,877 | 843,267 | 930,196 | 86,740 | 45,507 | 668,665 | 128,469 | 49,166 | 404,823 | 1,249,743 | 1,404,261 | 1,607,184 | 2,076,935 | 2,251,147 | 2,466,684 |
| 1988 | 26,762 | 24,023 | 499,271 | 131,925 | 27,547 | 649,846 | 711,598 | 783,380 | 29,257 | 81,517 | 909,270 | 176,564 | 115,867 | 611,568 | 1,750,813 | 1,940,973 | 2,164,376 | 2,451,167 | 2,653,503 | 2,888,079 |
| 1989 | 62,585 | 12,986 | 548,990 | 196,898 | 8,788 | 755,865 | 831,632 | 922,988 | 16,080 | 45,685 | 649,338 | 118,373 | 111,531 | 670,238 | 1,476,773 | 1,623,273 | 1,797,974 | 2,290,150 | 2,458,741 | 2,649,359 |
| 1990 | 59,283 | 9,696 | 491,249 | 163,341 | 19,562 | 678,399 | 745,070 | 824,726 | 26,758 | 42,019 | 407,593 | 84,815 | 92,224 | 320,661 | 895,049 | 984,174 | 1,088,153 | 1,617,130 | 1,732,713 | 1,857,975 |
| 1991 | 71,970 | 14,113 | 429,150 | 138,779 | 23,728 | 620,099 | 680,527 | 751,235 | 19,591 | 46,426 | 290,956 | 84,087 | 51,483 | 319,236 | 749,487 | 819,616 | 900,413 | 1,408,011 | 1,502,944 | 1,607,505 |
| 1992 | 95,447 | 26,524 | 361,201 | 171,414 | 25,868 | 628,101 | 684,749 | 748,026 | 35,702 | 52,994 | 421,994 | 87,963 | 104,483 | 465,556 | 1,079,082 | 1,181,790 | 1,300,906 | 1,746,352 | 1,869,066 | 1,999,952 |
| 1993 | 67,178 | 21,799 | 363,173 | 147,164 | 27,624 | 580,859 | 630,337 | 686,798 | 50,940 | 52,091 | 343,489 | 121,885 | 122,280 | 418,356 | 1,029,666 | 1,125,468 | 1,242,569 | 1,647,803 | 1,757,231 | 1,886,456 |
| 1994 | 26,700 | 6,978 | 491,611 | 173,473 | 21,046 | 652,710 | 722,561 | 805,454 | 40,192 | 42,727 | 439,899 | 135,754 | 83,916 | 445,734 | 1,096,872 | 1,203,956 | 1,328,518 | 1,797,513 | 1,930,740 | 2,075,239 |
| 1995 | 26,281 | 20,062 | 320,702 | 156,278 | 30,603 | 511,244 | 557,528 | 609,543 | 13,263 | 57,980 | 491,194 | 103,755 | 77,941 | 436,555 | 1,084,260 | 1,189,364 | 1,310,074 | 1,632,528 | 1,747,108 | 1,879,069 |
| 1996 | 60,980 | 10,704 | 244,566 | 212,063 | 18,902 | 505,646 | 550,611 | 602,463 | 16,546 | 50,166 | 457,514 | 76,770 | 80,471 | 313,171 | 908,870 | 1,003,220 | 1,113,825 | 1,447,693 | 1,555,958 | 1,674,972 |
| 1997 | 51,985 | 14,660 | 282,653 | 208,116 | 8,658 | 520,617 | 569,049 | 623,473 | 8,546 | 36,676 | 456,108 | 68,753 | 95,653 | 225,754 | 808,917 | 897,300 | 1,003,516 | 1,367,726 | 1,468,454 | 1,584,324 |
| 1998 | 60,032 | 24,861 | 368,874 | 228,187 | 7,626 | 632,076 | 693,432 | 760,066 | 16,675 | 50,116 | 478,388 | 75,528 | 207,890 | 307,340 | 1,045,374 | 1,146,650 | 1,264,395 | 1,720,966 | 1,841,845 | 1,977,840 |
| 1999 | 85,938 | 12,676 | 342,879 | 176,844 | 11,265 | 578,666 | 632,611 | 690,989 | 5,548 | 40,703 | 445,968 | 60,031 | 54,147 | 152,408 | 679,427 | 763,091 | 866,026 | 1,296,552 | 1,397,704 | 1,514,299 |
| 2000 | 90,688 | 13,312 | 563,974 | 192,546 | 22,397 | 809,781 | 887,085 | 976,135 | 14,430 | 36,156 | 619,127 | 91,357 | 78,688 | 297,578 | 1,027,924 | 1,145,490 | 1,289,953 | 1,890,113 | 2,035,397 | 2,199,454 |
| 2001 | 40,933 | 12,091 | 485,990 | 260,232 | 14,588 | 731,895 | 820,223 | 928,698 | 12,289 | 32,373 | 492,605 | 79,443 | 62,176 | 291,291 | 899,162 | 979,525 | 1,072,115 | 1,679,162 | 1,803,139 | 1,943,790 |
| 2002 | 28,712 | 20,965 | 297,616 | 236,586 | 14,841 | 535,897 | 603,162 | 695,441 | 28,004 | 40,512 | 430,206 | 75,245 | 123,106 | 234,744 | 868,351 | 944,371 | 1,029,549 | 1,444,025 | 1,551,632 | 1,673,183 |
| 2003 | 33,714 | 11,134 | 411,882 | 211,641 | 9,131 | 607,748 | 682,638 | 772,856 | 18,478 | 48,535 | 420,970 | 58,455 | 80,443 | 266,731 | 833,522 | 905,588 | 988,404 | 1,481,874 | 1,591,239 | 1,708,982 |
| 2004 | 13,153 | 30,066 | 250,021 | 147,794 | 7,866 | 407,168 | 451,894 | 509,646 | 22,005 | 48,423 | 310,109 | 104,246 | 71,642 | 316,409 | 812,020 | 888,680 | 979,005 | 1,251,643 | 1,343,117 | 1,447,666 |
| 2005 | 33,483 | 26,733 | 370,001 | 167,867 | 6,673 | 550,333 | 610,824 | 683,363 | 14,451 | 71,265 | 309,196 | 85,250 | 91,253 | 343,655 | 854,306 | 928,346 | 1,011,624 | 1,443,803 | 1,541,041 | 1,648,117 |
| 2006 | 63,470 | 28,229 | 300,080 | 203,587 | 8,065 | 546,745 | 607,301 | 684,571 | 20,242 | 50,314 | 237,036 | 84,587 | 58,133 | 332,716 | 725,998 | 797,247 | 881,220 | 1,309,591 | 1,408,289 | 1,519,392 |
| 2007 | 11,775 | 20,899 | 167,702 | 110,069 | 3,838 | 283,449 | 316,531 | 356,723 | 15,878 | 57,841 | 271,895 | 80,365 | 94,785 | 326,611 | 756,579 | 868,260 | 1,071,886 | 1,067,836 | 1,187,119 | 1,390,865 |
| 2008 | 12,126 | 19,114 | 210,252 | 114,542 | 4,970 | 325,710 | 364,142 | 410,152 | 15,565 | 70,276 | 267,404 | 78,737 | 56,640 | 281,738 | 681,494 | 795,623 | 993,227 | 1,040,162 | 1,162,613 | 1,362,225 |
| 2009 | 24,710 | 30,818 | 168,343 | 108,622 | 5,278 | 306,041 | 340,075 | 379,526 | 5,587 | 79,216 | 220,020 | 49,251 | 43,146 | 240,772 | 564,557 | 657,657 | 816,863 | 897,935 | 1,000,243 | 1,162,614 |
| 2010 | 23,071 | 24,696 | 249,554 | 123,776 | 8,845 | 390,238 | 433,217 | 482,094 | 18,932 | 80,946 | 279,976 | 98,082 | 39,540 | 440,563 | 843,942 | 995,448 | 1,225,239 | 1,271,659 | 1,430,029 | 1,660,304 |
| 2011 | 28,290 | 20,324 | 175,720 | 131,423 | 7,320 | 329,234 | 365,510 | 408,449 | 13,338 | 57,145 | 247,440 | 56,824 | 34,167 | 234,624 | 561,304 | 663,195 | 857,331 | 920,489 | 1,031,917 | 1,226,828 |
| 2012 | 61,214 | 8,316 | 195,651 | 152,538 | 7,896 | 386,746 | 429,062 | 482,818 | 11,579 | 35,677 | 252,034 | 35,002 | 42,991 | 328,367 | 609,083 | 736,068 | 937,401 | 1,031,442 | 1,168,728 | 1,376,404 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 yr Av. | 30,501 | 22,033 | 249,921 | 147,186 | 6,988 | 413,341 | 460,120 | 517,020 | 15,605 | 59,964 | 281,608 | 73,080 | 61,274 | 311,219 | 724,280 | 823,611 | 976,220 | 1,171,643 | 1,286,434 | 1,450,340 |



| Year | Northern Europe |  |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  |  |  | NEAC Area |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finland | Iceland | Norway | Russia | Sweden | Total |  |  | France | Iceland | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  |  | Total |  |  |
|  |  | N\&E |  |  |  | 5.0\% | 50.0\% | 95.0\% |  | S\&W |  |  |  |  | 5.0\% | 50.0\% | 95.0\% | 5.0\% | 50.0\% | 95.0\% |
| 1971 | 23,975 | 9,667 |  | 132,797 | 1,058 |  |  |  | 10,825 | 24,387 | 157,476 | 90,670 | 21,918 | 567,909 | 781,633 | 878,474 | 995,713 |  |  |  |
| 1972 | 37,414 | 15,070 |  | 134,665 | 744 |  |  |  | 21,535 | 37,462 | 169,264 | 150,146 | 19,122 | 731,177 | 1,006,433 | 1,137,847 | 1,292,236 |  |  |  |
| 1973 | 44,746 | 14,089 |  | 222,757 | 2,581 |  |  |  | 13,255 | 33,834 | 183,556 | 114,508 | 16,732 | 804,762 | 1,036,567 | 1,173,408 | 1,340,088 |  |  |  |
| 1974 | 66,536 | 13,378 |  | 209,676 | 1,661 |  |  |  | 6,210 | 29,215 | 206,560 | 84,432 | 18,287 | 569,348 | 815,078 | 921,572 | 1,046,745 |  |  |  |
| 1975 | 73,894 | 14,795 |  | 225,188 | 402 |  |  |  | 12,288 | 30,998 | 231,473 | 114,883 | 14,990 | 626,889 | 917,315 | 1,042,857 | 1,188,220 |  |  |  |
| 1976 | 60,884 | 12,156 |  | 194,865 | 1,210 |  |  |  | 8,986 | 26,790 | 160,133 | 60,268 | 10,441 | 390,985 | 585,755 | 664,745 | 752,483 |  |  |  |
| 1977 | 37,202 | 16,930 |  | 134,427 | 909 |  |  |  | 6,964 | 26,150 | 139,455 | 76,591 | 10,285 | 427,516 | 614,650 | 693,969 | 787,743 |  |  |  |
| 1978 | 23,670 | 21,839 |  | 115,986 | 694 |  |  |  | 7,117 | 33,829 | 120,591 | 64,405 | 13,401 | 533,374 | 686,419 | 778,610 | 888,193 |  |  |  |
| 1979 | 25,296 | 14,411 |  | 101,438 | 2,016 |  |  |  | 8,120 | 21,637 | 109,132 | 31,847 | 9,397 | 395,351 | 507,698 | 579,742 | 666,298 |  |  |  |
| 1980 | 26,463 | 20,064 |  | 168,999 | 3,529 |  |  |  | 17,055 | 30,387 | 119,991 | 103,798 | 11,907 | 483,147 | 689,980 | 774,062 | 875,850 |  |  |  |
| 1981 | 29,170 | 7,043 |  | 96,572 | 1,022 |  |  |  | 11,620 | 20,288 | 88,752 | 145,209 | 9,325 | 518,670 | 711,504 | 803,097 | 909,395 |  |  |  |
| 1982 | 38,211 | 8,078 |  | 85,292 | 3,685 |  |  |  | 7,255 | 14,322 | 51,687 | 56,375 | 13,509 | 417,891 | 501,807 | 563,981 | 646,372 |  |  |  |
| 1983 | 41,214 | 6,155 | 428,870 | 124,130 | 2,538 | 548,299 | 604,904 | 671,526 | 7,734 | 23,930 | 154,218 | 64,361 | 18,977 | 452,652 | 629,011 | 737,860 | 935,912 | 1,221,079 | 1,347,548 | 1,550,814 |
| 1984 | 39,177 | 7,945 | 439,174 | 123,807 | 3,553 | 558,996 | 616,043 | 680,524 | 12,551 | 20,252 | 76,464 | 51,557 | 7,446 | 377,198 | 491,776 | 548,844 | 616,701 | 1,084,028 | 1,167,298 | 1,258,442 |
| 1985 | 30,629 | 5,124 | 405,090 | 135,422 | 1,484 | 527,051 | 579,419 | 638,846 | 9,509 | 14,700 | 83,680 | 75,900 | 9,642 | 463,287 | 591,139 | 660,647 | 743,618 | 1,153,669 | 1,241,581 | 1,342,068 |
| 1986 | 26,762 | 13,921 | 486,655 | 133,699 | 1,434 | 601,567 | 664,851 | 736,928 | 9,740 | 12,270 | 94,635 | 103,167 | 10,865 | 593,310 | 738,552 | 830,181 | 940,805 | 1,381,701 | 1,497,727 | 1,627,887 |
| 1987 | 33,412 | 14,440 | 367,065 | 99,367 | 4,300 | 473,915 | 520,902 | 576,206 | 5,137 | 10,891 | 117,707 | 83,164 | 5,534 | 387,818 | 548,764 | 615,204 | 692,280 | 1,053,565 | 1,137,291 | 1,231,735 |
| 1988 | 21,716 | 9,313 | 306,088 | 99,762 | 4,168 | 404,042 | 442,174 | 485,823 | 14,243 | 12,440 | 84,933 | 108,256 | 15,662 | 601,153 | 751,315 | 842,747 | 951,924 | 1,185,239 | 1,286,478 | 1,401,418 |
| 1989 | 24,336 | 7,908 | 219,835 | 97,409 | 11,612 | 332,594 | 362,355 | 395,401 | 6,493 | 11,086 | 77,536 | 86,109 | 12,453 | 524,551 | 646,949 | 723,111 | 812,649 | 1,002,558 | 1,086,670 | 1,181,563 |
| 1990 | 30,495 | 8,321 | 260,039 | 124,651 | 7,430 | 397,817 | 431,976 | 473,069 | 6,693 | 10,980 | 37,168 | 106,903 | 11,327 | 439,482 | 548,664 | 617,445 | 696,197 | 972,668 | 1,051,480 | 1,140,157 |
| 1991 | 36,578 | 5,775 | 220,247 | 122,316 | 8,570 | 364,284 | 395,439 | 430,238 | 6,045 | 10,952 | 56,016 | 46,883 | 5,812 | 331,894 | 413,268 | 460,737 | 518,534 | 798,191 | 856,964 | 924,466 |
| 1992 | 39,145 | 8,616 | 239,432 | 116,453 | 11,019 | 383,835 | 416,153 | 453,166 | 7,573 | 12,352 | 42,996 | 35,696 | 13,315 | 443,558 | 498,337 | 557,735 | 632,141 | 905,669 | 975,689 | 1,056,781 |
| 1993 | 45,379 | 9,724 | 229,330 | 137,665 | 15,035 | 408,938 | 439,551 | 472,133 | 3,577 | 6,052 | 42,160 | 39,479 | 31,440 | 363,735 | 437,687 | 491,869 | 557,318 | 868,374 | 932,895 | 1,004,406 |
| 1994 | 37,650 | 8,266 | 224,469 | 121,800 | 11,040 | 374,343 | 405,608 | 440,343 | 7,633 | 9,812 | 67,511 | 55,496 | 11,042 | 442,520 | 533,612 | 598,330 | 673,389 | 931,782 | 1,004,888 | 1,087,161 |
| 1995 | 23,409 | 5,716 | 240,538 | 138,927 | 7,710 | 386,650 | 417,903 | 453,196 | 3,641 | 11,077 | 65,224 | 55,669 | 9,357 | 408,961 | 498,350 | 557,521 | 630,060 | 907,450 | 976,874 | 1,057,352 |
| 1996 | 20,557 | 7,538 | 241,608 | 104,418 | 9,898 | 355,607 | 386,113 | 419,181 | 6,454 | 7,136 | 43,565 | 57,461 | 10,217 | 312,220 | 391,999 | 441,959 | 500,891 | 769,587 | 828,749 | 896,242 |
| 1997 | 29,775 | 4,230 | 159,340 | 85,330 | 6,375 | 264,432 | 287,127 | 311,884 | 3,339 | 8,006 | 56,562 | 35,924 | 12,724 | 215,669 | 297,283 | 338,387 | 387,141 | 578,282 | 626,078 | 680,509 |
| 1998 | 25,336 | 6,170 | 191,388 | 105,483 | 4,729 | 309,799 | 334,886 | 363,004 | 2,807 | 4,965 | 32,885 | 23,214 | 17,483 | 227,752 | 279,119 | 311,973 | 351,857 | 604,496 | 647,897 | 694,973 |
| 1999 | 23,661 | 7,094 | 204,260 | 93,153 | 4,071 | 305,489 | 333,443 | 365,543 | 6,101 | 9,685 | 51,197 | 46,722 | 7,985 | 176,189 | 262,866 | 307,907 | 363,429 | 586,781 | 642,375 | 706,067 |
| 2000 | 52,693 | 4,162 | 282,034 | 162,211 | 8,870 | 474,562 | 512,599 | 555,451 | 4,248 | 2,631 | 63,883 | 48,181 | 10,625 | 225,003 | 319,536 | 360,664 | 410,717 | 817,784 | 874,550 | 936,620 |
| 2001 | 75,743 | 4,769 | 333,548 | 114,801 | 10,677 | 498,662 | 541,995 | 590,481 | 4,988 | 4,615 | 56,711 | 52,024 | 7,814 | 214,447 | 304,294 | 347,005 | 398,605 | 826,455 | 890,544 | 959,285 |
| 2002 | 60,451 | 4,508 | 289,370 | 125,075 | 7,827 | 449,251 | 489,248 | 534,577 | 4,619 | 4,993 | 65,579 | 46,828 | 9,267 | 175,450 | 275,832 | 313,746 | 358,652 | 747,963 | 803,516 | 867,904 |
| 2003 | 43,008 | 4,737 | 255,438 | 87,213 | 8,897 | 369,216 | 401,199 | 437,096 | 6,637 | 7,988 | 69,261 | 60,082 | 6,044 | 218,311 | 327,497 | 375,615 | 434,278 | 718,655 | 778,224 | 844,778 |
| 2004 | 20,646 | 4,651 | 231,629 | 67,252 | 6,491 | 303,460 | 331,788 | 365,008 | 12,358 | 6,456 | 38,000 | 51,498 | 5,398 | 282,011 | 352,156 | 402,610 | 462,471 | 678,305 | 735,308 | 801,673 |
| 2005 | 15,924 | 5,784 | 213,001 | 80,584 | 4,924 | 295,457 | 321,198 | 350,162 | 7,628 | 5,692 | 49,240 | 55,365 | 6,880 | 222,990 | 310,324 | 353,585 | 407,192 | 625,367 | 675,407 | 735,054 |
| 2006 | 27,779 | 5,534 | 270,422 | 77,243 | 4,915 | 355,219 | 387,146 | 422,943 | 7,730 | 4,726 | 35,763 | 50,108 | 4,396 | 231,175 | 296,198 | 342,604 | 397,358 | 672,739 | 730,160 | 795,478 |
| 2007 | 39,813 | 5,293 | 230,280 | 80,540 | 6,800 | 336,496 | 363,857 | 394,482 | 7,255 | 2,908 | 15,846 | 48,672 | 6,050 | 221,844 | 266,325 | 308,871 | 361,168 | 621,705 | 673,059 | 732,846 |
| 2008 | 37,915 | 6,823 | 265,318 | 125,920 | 9,768 | 409,815 | 448,680 | 492,255 | 8,047 | 3,321 | 23,998 | 53,727 | 3,648 | 249,639 | 300,300 | 349,472 | 411,157 | 734,516 | 799,568 | 872,637 |
| 2009 | 17,553 | 5,513 | 207,982 | 107,000 | 8,780 | 318,278 | 348,711 | 384,081 | 4,215 | 5,152 | 27,014 | 40,759 | 4,787 | 211,569 | 256,676 | 299,541 | 350,981 | 594,859 | 649,553 | 711,483 |
| 2010 | 27,891 | 7,833 | 229,078 | 132,362 | 10,755 | 374,259 | 410,331 | 449,227 | 3,557 | 10,664 | 17,417 | 60,625 | 4,370 | 278,651 | 324,595 | 383,248 | 457,997 | 724,596 | 795,077 | 877,266 |
| 2011 | 21,906 | 8,726 | 318,725 | 131,895 | 14,038 | 451,842 | 497,778 | 550,480 | 9,228 | 5,420 | 20,137 | 87,759 | 11,420 | 313,606 | 386,813 | 459,861 | 555,115 | 870,116 | 959,608 | 1,065,974 |
| 2012 | 28,360 | 4,267 | 279,528 | 65,039 | 16,436 | 357,995 | 395,360 | 438,426 | 7,210 | 4,649 | 20,726 | 70,760 | 14,206 | 260,554 | 324,260 | 388,630 | 469,655 | 708,546 | 785,663 | 875,645 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 yr Av. | 28,079 | 5,916 | 250,140 | 95,505 | 9,180 | 357,204 | 390,605 | 428,416 | 7,386 | 5,698 | 31,740 | 57,936 | 6,720 | 249,035 | 314,514 | 366,404 | 430,737 | 694,940 | 758,163 | 831,283 |

Table 3.3.5.3 Estimated pre-fishery abundance of MATURING 1SW salmon (potential 1SW returns) by NEAC country or region and year

| Year | Northern Europe |  |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  |  |  | NEAC Area |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finland | Iceland | Norway | Russia | Sweden | Total |  |  | France | $\begin{array}{\|c\|} \hline \text { Iceland } \\ \hline \text { S\&W } \\ \hline \end{array}$ | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  |  | Total |  |  |
|  |  | N\&E |  |  |  | 5.0\% | 50.0\% | 95.0\% |  |  |  |  |  |  | 5.0\% | 50.0\% | 95.0\% | 5.0\% | 50.0\% | 95.0\% |
| 1971 | 33,545 | 12,094 |  | NA | 22,279 |  |  |  | 63,680 | 79,581 | 1,344,969 | 105,202 | 231,178 | 783,005 | 2,272,283 | 2,622,949 | 3,045,619 |  |  |  |
| 1972 | 51,958 | 11,092 |  | 151,275 | 17,659 |  |  |  | 127,611 | 64,806 | 1,427,891 | 101,388 | 202,207 | 683,229 | 2,273,962 | 2,628,999 | 3,086,322 |  |  |  |
| 1973 | 47,287 | 13,251 |  | 223,861 | 21,931 |  |  |  | 76,976 | 69,272 | 1,561,580 | 119,489 | 176,586 | 818,985 | 2,443,862 | 2,836,888 | 3,323,639 |  |  |  |
| 1974 | 93,670 | 13,254 |  | 222,530 | 31,449 |  |  |  | 36,228 | 49,587 | 1,774,547 | 149,571 | 193,486 | 781,025 | 2,573,849 | 3,001,384 | 3,531,743 |  |  |  |
| 1975 | 65,162 | 16,083 |  | 340,869 | 33,891 |  |  |  | 72,113 | 76,401 | 1,963,846 | 152,978 | 158,674 | 636,008 | 2,624,550 | 3,073,860 | 3,654,325 |  |  |  |
| 1976 | 44,934 | 16,150 |  | 237,065 | 19,205 |  |  |  | 65,917 | 60,402 | 1,335,632 | 102,772 | 110,520 | 548,068 | 1,915,099 | 2,233,631 | 2,642,125 |  |  |  |
| 1977 | 23,186 | 22,416 |  | 151,521 | 9,135 |  |  |  | 51,065 | 61,850 | 1,147,164 | 116,360 | 108,769 | 571,532 | 1,789,105 | 2,069,616 | 2,431,613 |  |  |  |
| 1978 | 31,303 | 22,772 |  | 153,140 | 10,356 |  |  |  | 52,340 | 81,015 | 1,007,321 | 133,148 | 141,662 | 654,926 | 1,814,605 | 2,082,162 | 2,415,707 |  |  |  |
| 1979 | 36,735 | 21,878 |  | 212,912 | 10,962 |  |  |  | 60,327 | 74,966 | 930,902 | 126,790 | 99,438 | 539,777 | 1,601,026 | 1,845,464 | 2,137,336 |  |  |  |
| 1980 | 17,122 | 3,512 |  | 151,679 | 13,945 |  |  |  | 124,706 | 34,327 | 707,742 | 119,489 | 126,224 | 337,655 | 1,273,286 | 1,465,250 | 1,698,014 |  |  |  |
| 1981 | 26,726 | 17,370 |  | 127,279 | 25,409 |  |  |  | 98,494 | 44,505 | 374,509 | 126,153 | 99,707 | 418,108 | 1,047,724 | 1,174,504 | 1,317,751 |  |  |  |
| 1982 | 8,587 | 8,148 |  | 111,063 | 22,230 |  |  |  | 61,959 | 45,550 | 771,386 | 106,657 | 143,283 | 597,748 | 1,539,868 | 1,737,600 | 1,963,855 |  |  |  |
| 1983 | 37,726 | 11,932 | 895,529 | 184,774 | 29,576 | 1,015,958 | 1,162,491 | 1,332,641 | 65,931 | 57,558 | 1,360,399 | 156,317 | 200,326 | 611,390 | 2,161,066 | 2,464,818 | 2,824,543 | 3,247,194 | 3,628,731 | 4,079,555 |
| 1984 | 41,046 | 4,333 | 928,372 | 196,771 | 41,094 | 1,061,649 | 1,215,676 | 1,397,807 | 107,956 | 35,269 | 711,747 | 136,112 | 78,831 | 642,945 | 1,532,145 | 1,726,587 | 1,958,795 | 2,656,584 | 2,946,593 | 3,277,789 |
| 1985 | 61,513 | 28,974 | 944,167 | 269,475 | 48,997 | 1,197,824 | 1,356,907 | 1,547,129 | 40,353 | 56,851 | 1,180,700 | 136,450 | 102,145 | 532,277 | 1,800,918 | 2,057,295 | 2,367,564 | 3,067,818 | 3,419,983 | 3,831,519 |
| 1986 | 56,721 | 36,255 | 825,263 | 231,437 | 51,736 | 1,071,115 | 1,207,354 | 1,370,843 | 62,033 | 93,709 | 1,320,206 | 158,471 | 115,146 | 660,789 | 2,136,004 | 2,439,336 | 2,804,964 | 3,273,439 | 3,651,207 | 4,087,224 |
| 1987 | 71,562 | 21,293 | 692,725 | 246,317 | 41,952 | 955,930 | 1,079,081 | 1,221,569 | 110,205 | 58,145 | 850,014 | 163,833 | 62,888 | 510,299 | 1,554,607 | 1,786,541 | 2,081,943 | 2,569,259 | 2,871,309 | 3,227,374 |
| 1988 | 34,784 | 30,720 | 637,761 | 170,207 | 35,176 | 810,824 | 911,189 | 1,031,715 | 37,550 | 104,061 | 1,155,072 | 224,673 | 148,019 | 771,888 | 2,164,091 | 2,461,118 | 2,818,590 | 3,025,291 | 3,376,883 | 3,784,649 |
| 1989 | 80,278 | 16,683 | 700,565 | 252,459 | 11,370 | 941,715 | 1,064,071 | 1,212,042 | 20,724 | 58,322 | 826,033 | 151,186 | 142,436 | 845,082 | 1,827,382 | 2,058,520 | 2,338,339 | 2,822,039 | 3,129,731 | 3,475,861 |
| 1990 | 75,789 | 12,426 | 626,698 | 208,281 | 24,966 | 841,732 | 950,600 | 1,079,351 | 34,157 | 53,508 | 517,547 | 108,018 | 117,378 | 404,572 | 1,105,859 | 1,249,138 | 1,422,368 | 1,988,195 | 2,202,641 | 2,448,820 |
| 1991 | 91,730 | 17,986 | 545,306 | 178,085 | 30,272 | 770,204 | 868,399 | 983,364 | 25,074 | 59,177 | 370,250 | 106,811 | 65,611 | 402,321 | 925,226 | 1,039,975 | 1,176,317 | 1,733,913 | 1,910,813 | 2,114,740 |
| 1992 | 121,328 | 33,759 | 460,052 | 218,914 | 32,841 | 776,117 | 872,214 | 978,937 | 45,298 | 67,442 | 535,873 | 111,592 | 132,729 | 586,452 | 1,329,785 | 1,498,163 | 1,689,902 | 2,144,276 | 2,370,101 | 2,619,193 |
| 1993 | 85,557 | 27,701 | 461,831 | 188,279 | 35,297 | 717,501 | 802,746 | 903,033 | 64,900 | 66,333 | 436,160 | 155,024 | 155,518 | 526,015 | 1,269,074 | 1,427,038 | 1,617,808 | 2,020,962 | 2,231,558 | 2,474,740 |
| 1994 | 34,122 | 8,912 | 624,066 | 222,790 | 26,848 | 810,072 | 922,727 | 1,054,550 | 51,268 | 54,324 | 559,600 | 172,792 | 106,748 | 561,071 | 1,351,841 | 1,527,357 | 1,727,595 | 2,211,472 | 2,451,959 | 2,720,742 |
| 1995 | 33,509 | 25,530 | 407,767 | 200,378 | 38,986 | 631,453 | 710,958 | 800,232 | 16,863 | 73,791 | 623,375 | 132,015 | 99,017 | 549,907 | 1,335,827 | 1,505,330 | 1,703,831 | 2,000,345 | 2,218,771 | 2,466,790 |
| 1996 | 77,670 | 13,641 | 311,239 | 272,052 | 24,009 | 624,990 | 702,275 | 791,213 | 21,006 | 63,712 | 581,856 | 97,759 | 102,430 | 394,827 | 1,120,850 | 1,271,152 | 1,445,824 | 1,778,707 | 1,977,088 | 2,195,101 |
| 1997 | 66,109 | 18,651 | 358,917 | 266,938 | 11,016 | 644,004 | 725,216 | 818,928 | 10,832 | 46,640 | 579,835 | 87,484 | 121,693 | 284,381 | 999,263 | 1,139,233 | 1,303,129 | 1,677,495 | 1,865,951 | 2,077,904 |
| 1998 | 76,404 | 31,578 | 468,531 | 292,477 | 9,680 | 784,344 | 883,567 | 998,879 | 21,282 | 63,713 | 607,831 | 96,501 | 264,724 | 386,983 | 1,288,734 | 1,451,796 | 1,644,930 | 2,115,731 | 2,339,437 | 2,596,095 |
| 1999 | 109,190 | 16,102 | 435,452 | 226,187 | 14,291 | 716,860 | 804,074 | 905,993 | 7,061 | 51,717 | 565,615 | 76,285 | 68,754 | 191,831 | 841,670 | 968,857 | 1,119,872 | 1,596,492 | 1,774,687 | 1,979,683 |
| 2000 | 115,262 | 16,967 | 716,003 | 246,935 | 28,465 | 1,002,249 | 1,129,735 | 1,281,016 | 18,380 | 45,931 | 786,005 | 116,286 | 99,981 | 374,316 | 1,270,571 | 1,451,111 | 1,670,686 | 2,324,523 | 2,587,684 | 2,880,808 |
| 2001 | 52,170 | 15,380 | 617,427 | 333,014 | 18,526 | 909,342 | 1,045,340 | 1,210,145 | 15,655 | 41,148 | 626,389 | 101,052 | 79,036 | 366,463 | 1,105,307 | 1,240,359 | 1,398,736 | 2,062,515 | 2,288,219 | 2,550,331 |
| 2002 | 36,524 | 26,677 | 378,003 | 303,392 | 18,892 | 666,778 | 769,689 | 906,536 | 35,589 | 51,487 | 547,562 | 95,518 | 156,360 | 295,512 | 1,067,617 | 1,196,382 | 1,346,033 | 1,772,980 | 1,971,660 | 2,197,600 |
| 2003 | 42,880 | 14,173 | 523,821 | 270,156 | 11,612 | 755,413 | 869,085 | 1,005,513 | 23,495 | 61,710 | 535,916 | 74,284 | 102,298 | 335,746 | 1,023,088 | 1,147,994 | 1,288,631 | 1,817,428 | 2,020,808 | 2,243,810 |
| 2004 | 16,683 | 38,241 | 317,729 | 188,898 | 10,025 | 504,851 | 575,963 | 664,574 | 28,096 | 61,493 | 394,457 | 132,837 | 91,130 | 398,857 | 999,534 | 1,124,900 | 1,273,737 | 1,536,578 | 1,704,229 | 1,895,595 |
| 2005 | 42,502 | 34,068 | 470,470 | 215,551 | 8,490 | 682,656 | 777,704 | 892,001 | 18,411 | 90,741 | 394,244 | 108,382 | 116,004 | 432,647 | 1,051,910 | 1,174,652 | 1,319,335 | 1,772,082 | 1,955,792 | 2,169,424 |
| 2006 | 80,805 | 35,892 | 381,191 | 259,942 | 10,254 | 678,102 | 774,458 | 893,536 | 25,756 | 63,982 | 301,754 | 107,291 | 74,019 | 419,235 | 895,161 | 1,008,784 | 1,143,607 | 1,611,407 | 1,786,903 | 1,990,255 |
| 2007 | 14,960 | 26,606 | 213,046 | 140,489 | 4,875 | 352,004 | 402,926 | 465,545 | 20,262 | 73,602 | 345,331 | 101,960 | 120,318 | 411,514 | 938,809 | 1,102,597 | 1,369,678 | 1,322,194 | 1,511,251 | 1,790,576 |
| 2008 | 15,396 | 24,322 | 266,897 | 146,326 | 6,306 | 404,024 | 463,352 | 535,297 | 19,845 | 89,279 | 340,347 | 100,281 | 72,051 | 355,559 | 845,070 | 1,010,048 | 1,269,876 | 1,284,160 | 1,479,699 | 1,752,742 |
| 2009 | 31,416 | 39,223 | 214,053 | 137,400 | 6,724 | 378,272 | 431,555 | 493,193 | 7,123 | 100,756 | 280,015 | 62,853 | 54,682 | 303,320 | 699,644 | 834,237 | 1,046,344 | 1,108,830 | 1,270,789 | 1,497,160 |
| 2010 | 29,320 | 31,442 | 317,260 | 156,751 | 11,223 | 482,475 | 550,298 | 629,039 | 24,041 | 102,983 | 357,136 | 124,852 | 50,319 | 554,961 | 1,048,289 | 1,259,296 | 1,567,547 | 1,569,645 | 1,814,632 | 2,140,422 |
| 2011 | 35,990 | 25,821 | 223,284 | 166,719 | 9,308 | 407,397 | 464,263 | 531,184 | 16,960 | 72,672 | 314,478 | 72,330 | 43,586 | 295,744 | 698,598 | 843,064 | 1,095,288 | 1,136,871 | 1,313,648 | 1,578,571 |
| 2012 | 78,015 | 10,587 | 248,611 | 194,063 | 10,038 | 478,316 | 546,557 | 630,002 | 14,704 | 45,205 | 320,690 | 44,519 | 54,498 | 414,460 | 758,880 | 932,133 | 1,205,844 | 1,275,909 | 1,485,806 | 1,775,052 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 10yr Av. | 38,797 | 28,038 | 317,636 | 187,630 | 8,885 | 512,351 | 585,616 | 673,988 | 19,869 | 76,242 | 358,437 | 92,959 | 77,891 | 392,204 | 895,898 | 1,043,771 | 1,257,989 | 1,443,510 | 1,634,356 | 1,883,361 |



| Year | Northern Europe |  |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  |  |  | NEAC Area |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finland | Iceland | Norway | Russia | Sweden | Total |  |  | France | Iceland | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  |  | Total |  |  |
|  |  | N\&E |  |  |  | 5.0\% | 50.0\% | 95.0\% |  | S\&W |  |  |  |  | 5.0\% | 50.0\% | 95.0\% | 5.0\% | 50.0\% | 95.0\% |
| 1971 | 71,642 | 27,304 |  | 262,326 | 5,888 |  |  |  | 54,723 | 63,966 | 385,078 | 361,141 | 34,188 | 1,706,302 | 2,211,494 | 2,621,181 | 3,114,208 |  |  |  |
| 1972 | 85,035 | 25,819 |  | 416,715 | 8,857 |  |  |  | 36,778 | 57,936 | 387,828 | 277,581 | 30,494 | 1,742,261 | 2,133,614 | 2,545,763 | 3,038,110 |  |  |  |
| 1973 | 118,310 | 23,955 |  | 387,219 | 5,957 |  |  |  | 20,338 | 49,779 | 400,882 | 200,820 | 32,340 | 1,224,912 | 1,626,892 | 1,940,617 | 2,323,575 |  |  |  |
| 1974 | 132,943 | 26,771 |  | 420,624 | 4,625 |  |  |  | 32,178 | 53,173 | 452,287 | 263,424 | 27,375 | 1,370,700 | 1,851,248 | 2,211,672 | 2,655,530 |  |  |  |
| 1975 | 107,882 | 21,781 |  | 358,508 | 5,155 |  |  |  | 27,618 | 45,725 | 337,893 | 173,512 | 18,983 | 979,657 | 1,350,596 | 1,590,410 | 1,875,349 |  |  |  |
| 1976 | 66,435 | 29,277 |  | 248,124 | 3,663 |  |  |  | 19,854 | 44,392 | 279,324 | 177,873 | 18,219 | 934,438 | 1,244,998 | 1,483,213 | 1,771,116 |  |  |  |
| 1977 | 42,886 | 37,239 |  | 211,108 | 2,856 |  |  |  | 18,774 | 56,978 | 239,781 | 149,041 | 23,130 | 1,073,242 | 1,307,737 | 1,570,450 | 1,893,733 |  |  |  |
| 1978 | 47,692 | 25,304 |  | 195,502 | 5,638 |  |  |  | 19,415 | 36,910 | 214,905 | 87,133 | 17,124 | 825,847 | 1,005,510 | 1,207,628 | 1,461,069 |  |  |  |
| 1979 | 58,678 | 36,673 |  | 338,621 | 11,067 |  |  |  | 36,692 | 52,685 | 249,761 | 226,476 | 23,678 | 1,073,123 | 1,404,331 | 1,670,122 | 2,001,610 |  |  |  |
| 1980 | 71,461 | 16,735 |  | 240,497 | 9,526 |  |  |  | 27,987 | 37,094 | 201,349 | 299,904 | 21,754 | 1,177,592 | 1,489,098 | 1,775,952 | 2,121,774 |  |  |  |
| 1981 | 85,067 | 18,153 |  | 216,893 | 13,373 |  |  |  | 19,719 | 26,898 | 134,290 | 141,749 | 28,298 | 977,060 | 1,124,377 | 1,332,259 | 1,588,506 |  |  |  |
| 1982 | 85,464 | 13,918 | 812,981 | 269,475 | 9,875 | 1,001,338 | 1,195,711 | 1,427,316 | 18,962 | 42,276 | 295,953 | 146,278 | 36,096 | 975,721 | 1,264,546 | 1,548,446 | 1,962,953 | 2,307,851 | 2,752,651 | 3,313,079 |
| 1983 | 76,161 | 15,543 | 794,064 | 249,544 | 9,455 | 959,603 | 1,147,564 | 1,371,177 | 24,172 | 35,239 | 147,570 | 107,252 | 15,254 | 755,358 | 910,515 | 1,089,767 | 1,307,372 | 1,898,937 | 2,241,462 | 2,641,896 |
| 1984 | 62,459 | 11,001 | 740,397 | 270,632 | 6,194 | 918,590 | 1,091,964 | 1,307,897 | 18,597 | 26,118 | 157,438 | 147,136 | 19,156 | 894,363 | 1,052,266 | 1,267,670 | 1,528,349 | 2,000,229 | 2,363,531 | 2,798,288 |
| 1985 | 58,157 | 26,208 | 890,452 | 274,491 | 7,130 | 1,057,562 | 1,259,598 | 1,511,373 | 22,475 | 22,396 | 195,512 | 213,690 | 21,709 | 1,204,730 | 1,409,384 | 1,687,721 | 2,031,303 | 2,501,965 | 2,952,064 | 3,489,227 |
| 1986 | 68,501 | 26,896 | 688,984 | 212,506 | 11,706 | 851,138 | 1,012,488 | 1,210,956 | 15,508 | 19,974 | 237,376 | 182,458 | 12,672 | 869,565 | 1,135,596 | 1,344,232 | 1,602,699 | 2,015,133 | 2,357,233 | 2,778,674 |
| 1987 | 42,626 | 16,961 | 549,394 | 195,616 | 9,348 | 683,811 | 816,017 | 975,578 | 28,459 | 21,621 | 168,330 | 212,541 | 27,813 | 1,150,895 | 1,347,670 | 1,617,210 | 1,952,278 | 2,060,212 | 2,432,257 | 2,891,393 |
| 1988 | 48,770 | 14,943 | 415,549 | 195,507 | 22,286 | 589,629 | 699,000 | 831,758 | 16,857 | 19,666 | 164,146 | 182,337 | 22,867 | 1,058,317 | 1,232,015 | 1,467,945 | 1,762,075 | 1,843,577 | 2,167,454 | 2,566,086 |
| 1989 | 56,815 | 15,218 | 469,182 | 234,577 | 14,487 | 666,103 | 792,189 | 945,774 | 13,613 | 19,170 | 76,935 | 198,126 | 20,488 | 823,094 | 959,165 | 1,157,089 | 1,408,529 | 1,650,726 | 1,952,323 | 2,321,702 |
| 1990 | 64,820 | 10,428 | 387,579 | 223,135 | 15,551 | 591,818 | 704,006 | 840,655 | 11,452 | 18,758 | 101,354 | 88,323 | 10,661 | 605,171 | 695,371 | 840,715 | 1,020,033 | 1,306,448 | 1,546,201 | 1,834,314 |
| 1991 | 67,178 | 14,747 | 410,256 | 204,959 | 19,206 | 602,102 | 719,086 | 859,225 | 14,772 | 20,931 | 83,728 | 73,475 | 22,650 | 805,022 | 845,534 | 1,024,021 | 1,243,881 | 1,469,088 | 1,740,799 | 2,074,874 |
| 1992 | 77,287 | 16,609 | 392,064 | 241,168 | 25,732 | 637,603 | 755,187 | 900,180 | 7,533 | 10,336 | 79,188 | 76,858 | 52,763 | 657,438 | 735,268 | 891,669 | 1,086,335 | 1,391,930 | 1,648,193 | 1,958,535 |
| 1993 | 64,793 | 14,140 | 383,175 | 217,553 | 18,974 | 586,875 | 701,254 | 839,079 | 12,964 | 16,602 | 114,395 | 97,200 | 18,868 | 758,631 | 839,297 | 1,022,546 | 1,256,958 | 1,448,657 | 1,727,111 | 2,068,673 |
| 1994 | 41,279 | 10,006 | 413,071 | 246,055 | 13,582 | 608,134 | 725,029 | 866,981 | 6,359 | 18,735 | 110,967 | 97,693 | 16,123 | 703,821 | 787,551 | 959,806 | 1,180,353 | 1,417,034 | 1,685,894 | 2,018,806 |
| 1995 | 35,983 | 12,912 | 411,427 | 186,380 | 17,124 | 558,555 | 666,407 | 796,101 | 11,418 | 12,147 | 76,565 | 102,859 | 17,599 | 548,870 | 635,770 | 774,765 | 950,737 | 1,213,779 | 1,443,386 | 1,719,943 |
| 1996 | 49,825 | 7,068 | 265,729 | 147,792 | 10,700 | 404,358 | 484,018 | 578,717 | 5,922 | 13,377 | 96,028 | 63,737 | 21,244 | 371,845 | 472,352 | 581,321 | 719,143 | 892,495 | 1,065,987 | 1,274,632 |
| 1997 | 42,438 | 10,321 | 318,594 | 181,145 | 7,944 | 471,631 | 562,554 | 676,343 | 4,851 | 8,299 | 55,594 | 41,074 | 29,267 | 388,717 | 434,671 | 532,048 | 651,437 | 923,280 | 1,095,776 | 1,304,774 |
| 1998 | 39,433 | 11,824 | 340,502 | 161,922 | 6,800 | 468,296 | 561,659 | 675,349 | 10,256 | 16,148 | 86,031 | 80,846 | 13,356 | 298,758 | 413,789 | 520,153 | 659,954 | 900,798 | 1,085,010 | 1,308,535 |
| 1999 | 87,876 | 6,966 | 468,876 | 281,128 | 14,865 | 723,903 | 861,467 | 1,031,507 | 7,140 | 4,401 | 107,100 | 83,707 | 17,769 | 381,360 | 496,637 | 608,684 | 750,356 | 1,240,744 | 1,471,499 | 1,754,641 |
| 2000 | 126,391 | 7,975 | 554,777 | 199,715 | 17,966 | 760,963 | 909,479 | 1,094,119 | 8,768 | 7,719 | 97,477 | 91,781 | 13,010 | 372,226 | 486,706 | 599,678 | 742,425 | 1,272,210 | 1,510,129 | 1,804,231 |
| 2001 | 101,035 | 7,539 | 481,777 | 217,989 | 13,107 | 685,566 | 825,068 | 987,963 | 7,877 | 8,358 | 110,813 | 81,776 | 15,493 | 300,306 | 433,673 | 535,062 | 659,579 | 1,141,171 | 1,358,997 | 1,620,059 |
| 2002 | 71,841 | 7,918 | 425,506 | 152,242 | 14,875 | 563,882 | 674,173 | 807,421 | 11,360 | 13,304 | 117,115 | 105,170 | 10,119 | 372,391 | 517,849 | 641,155 | 797,158 | 1,103,019 | 1,314,931 | 1,574,677 |
| 2003 | 34,449 | 7,783 | 385,145 | 117,912 | 10,852 | 464,324 | 557,193 | 672,499 | 20,831 | 10,778 | 63,783 | 88,835 | 9,040 | 477,156 | 549,773 | 679,614 | 844,352 | 1,033,883 | 1,239,768 | 1,489,625 |
| 2004 | 26,578 | 9,654 | 355,166 | 140,680 | 8,232 | 452,535 | 541,595 | 648,474 | 12,813 | 9,480 | 82,605 | 95,977 | 11,469 | 375,695 | 484,254 | 598,027 | 743,554 | 955,969 | 1,140,240 | 1,366,181 |
| 2005 | 46,428 | 9,252 | 450,428 | 133,960 | 8,229 | 542,061 | 650,036 | 780,406 | 12,976 | 7,894 | 60,085 | 86,948 | 7,316 | 391,310 | 463,025 | 577,776 | 723,003 | 1,026,287 | 1,230,183 | 1,477,048 |
| 2006 | 66,497 | 8,851 | 382,963 | 137,972 | 11,371 | 512,433 | 608,924 | 727,538 | 12,218 | 4,849 | 27,180 | 84,402 | 10,074 | 375,941 | 418,353 | 523,071 | 657,016 | 947,329 | 1,134,508 | 1,358,597 |
| 2007 | 63,259 | 11,413 | 441,948 | 220,260 | 16,366 | 626,290 | 755,533 | 909,462 | 13,577 | 5,549 | 40,694 | 92,901 | 6,096 | 422,123 | 471,206 | 590,627 | 744,807 | 1,127,436 | 1,349,848 | 1,622,784 |
| 2008 | 29,447 | 9,211 | 346,777 | 185,339 | 14,631 | 487,426 | 587,723 | 709,347 | 7,124 | 8,603 | 45,623 | 70,653 | 7,943 | 357,835 | 402,914 | 505,209 | 638,388 | 911,126 | 1,094,230 | 1,320,006 |
| 2009 | 46,446 | 13,084 | 380,639 | 230,462 | 17,946 | 576,274 | 690,764 | 834,884 | 5,976 | 17,768 | 29,332 | 104,232 | 7,329 | 470,984 | 510,450 | 646,728 | 827,886 | 1,112,264 | 1,342,629 | 1,630,912 |
| 2010 | 36,678 | 14,607 | 530,311 | 229,083 | 23,359 | 692,941 | 837,350 | 1,013,814 | 15,549 | 9,042 | 34,314 | 152,102 | 19,148 | 534,375 | 610,891 | 780,108 | 1,001,921 | 1,338,563 | 1,618,983 | 1,960,017 |
| 2011 | 47,311 | 7,136 | 464,174 | 112,495 | 27,550 | 548,660 | 660,922 | 801,781 | 12,092 | 7,747 | 34,862 | 122,047 | 23,715 | 439,323 | 512,724 | 657,023 | 841,650 | 1,092,602 | 1,319,950 | 1,602,844 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 yr Av . | 46,893 | 9,891 | 416,306 | 166,040 | 15,341 | 546,682 | 656,421 | 790,562 | 12,452 | 9,501 | 53,559 | 100,327 | 11,225 | 421,713 | 494,144 | 619,934 | 781,974 | 1,064,848 | 1,278,527 | 1,540,269 |



|  | Northern Europe |  |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  |  |  | NEAC Area |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Finland | Iceland | Norway | Russia | Sweden | Total |  |  | France | Iceland | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  |  | Total |  |  |
|  |  | N\&E |  |  |  | 5.0\% | 50.0\% | 95.0\% |  | S\&W |  |  |  |  | 5.0\% | 50.0\% | 95.0\% | 5.0\% | 50.0\% | 95.0\% |
| 1971 | 12,950 | 4,697 |  |  | 8,182 |  |  |  | 48,271 | 31,265 | 394,571 | 35,189 | 36,435 | 213,877 | 583,042 | 770,899 | 1,023,905 |  |  |  |
| 1972 | 20,180 | 4,312 |  | 72,085 | 6,501 |  |  |  | 96,820 | 25,444 | 419,154 | 38,701 | 31,780 | 167,785 | 593,796 | 798,737 | 1,065,598 |  |  |  |
| 1973 | 18,380 | 5,174 |  | 78,472 | 8,112 |  |  |  | 58,343 | 27,153 | 457,879 | 46,364 | 27,801 | 204,435 | 616,949 | 835,701 | 1,126,239 |  |  |  |
| 1974 | 36,472 | 5,117 |  | 94,029 | 11,605 |  |  |  | 27,258 | 19,418 | 516,786 | 58,058 | 30,393 | 172,975 | 601,687 | 834,826 | 1,162,005 |  |  |  |
| 1975 | 25,364 | 6,287 |  | 111,983 | 12,508 |  |  |  | 54,610 | 29,875 | 575,301 | 60,084 | 24,912 | 154,958 | 658,403 | 910,056 | 1,267,006 |  |  |  |
| 1976 | 17,536 | 6,285 |  | 109,539 | 7,085 |  |  |  | 49,904 | 23,642 | 391,847 | 39,711 | 17,360 | 159,403 | 511,829 | 691,112 | 938,437 |  |  |  |
| 1977 | 8,957 | 8,764 |  | 74,428 | 3,356 |  |  |  | 38,669 | 24,222 | 336,520 | 45,045 | 17,079 | 140,180 | 453,352 | 608,590 | 825,532 |  |  |  |
| 1978 | 12,196 | 8,896 |  | 58,932 | 3,840 |  |  |  | 39,569 | 31,802 | 294,294 | 52,895 | 22,296 | 188,282 | 493,734 | 637,230 | 830,287 |  |  |  |
| 1979 | 14,256 | 8,544 |  | 75,257 | 4,006 |  |  |  | 45,644 | 29,492 | 275,131 | 51,656 | 15,584 | 124,529 | 417,969 | 551,397 | 725,077 |  |  |  |
| 1980 | 6,395 | 1,289 |  | 73,550 | 5,055 |  |  |  | 94,446 | 13,325 | 207,380 | 48,549 | 19,805 | 81,850 | 368,313 | 478,467 | 616,524 |  |  |  |
| 1981 | 9,851 | 6,655 |  | 53,779 | 9,208 |  |  |  | 74,255 | 17,290 | 70,030 | 51,796 | 15,529 | 98,594 | 270,162 | 337,926 | 414,185 |  |  |  |
| 1982 | 2,898 | 3,076 |  | 49,742 | 8,074 |  |  |  | 46,700 | 17,692 | 169,148 | 43,511 | 22,486 | 170,591 | 380,306 | 479,978 | 596,421 |  |  |  |
| 1983 | 14,315 | 4,565 | 160,753 | 64,867 | 10,805 | 202,548 | 256,512 | 320,281 | 49,645 | 22,385 | 361,624 | 64,250 | 31,411 | 149,393 | 536,788 | 687,624 | 874,907 | 782,264 | 947,755 | 1,145,367 |
| 1984 | 15,907 | 1,646 | 163,518 | 80,587 | 15,101 | 219,453 | 278,397 | 345,663 | 81,704 | 13,786 | 197,172 | 56,079 | 12,317 | 188,972 | 455,513 | 562,610 | 682,582 | 718,770 | 843,566 | 977,666 |
| 1985 | 23,996 | 11,398 | 171,329 | 92,776 | 18,210 | 259,749 | 320,464 | 387,124 | 30,588 | 22,354 | 236,835 | 55,972 | 16,006 | 178,938 | 417,703 | 547,097 | 709,642 | 723,947 | 870,848 | 1,043,099 |
| 1986 | 21,911 | 14,147 | 151,941 | 102,392 | 19,068 | 257,799 | 311,991 | 372,218 | 44,980 | 36,718 | 324,800 | 65,747 | 17,949 | 223,789 | 574,933 | 735,174 | 932,409 | 879,787 | 1,050,053 | 1,251,489 |
| 1987 | 27,783 | 8,269 | 127,306 | 95,726 | 15,490 | 231,086 | 276,901 | 328,759 | 80,727 | 22,766 | 199,545 | 68,965 | 15,265 | 169,575 | 453,027 | 582,102 | 756,577 | 721,531 | 860,613 | 1,040,716 |
| 1988 | 13,326 | 11,995 | 117,323 | 86,823 | 13,059 | 205,030 | 245,206 | 290,318 | 27,194 | 40,655 | 343,644 | 95,542 | 41,242 | 383,623 | 796,006 | 950,741 | 1,132,735 | 1,036,848 | 1,197,204 | 1,382,397 |
| 1989 | 25,063 | 6,507 | 184,803 | 96,355 | 4,137 | 267,822 | 318,517 | 380,243 | 14,956 | 22,814 | 220,991 | 64,333 | 12,382 | 440,331 | 664,742 | 789,281 | 933,686 | 974,083 | 1,110,832 | 1,263,177 |
| 1990 | 23,749 | 4,846 | 164,869 | 97,173 | 10,736 | 258,404 | 303,256 | 357,070 | 24,872 | 20,997 | 158,878 | 46,281 | 35,052 | 197,476 | 420,682 | 495,236 | 581,340 | 711,125 | 801,305 | 900,583 |
| 1991 | 28,693 | 7,064 | 144,138 | 83,245 | 13,094 | 237,648 | 278,544 | 325,700 | 18,229 | 23,268 | 117,542 | 46,831 | 18,296 | 215,054 | 387,452 | 448,830 | 518,209 | 652,895 | 728,593 | 810,750 |
| 1992 | 37,931 | 13,247 | 121,535 | 116,239 | 14,155 | 267,267 | 306,627 | 349,249 | 33,212 | 26,469 | 158,716 | 49,429 | 46,063 | 333,392 | 573,212 | 661,567 | 765,112 | 871,322 | 969,598 | 1,078,476 |
| 1993 | 26,893 | 10,899 | 120,760 | 113,882 | 15,133 | 252,912 | 290,062 | 331,683 | 47,359 | 26,101 | 141,025 | 72,078 | 72,250 | 275,490 | 565,873 | 651,946 | 758,546 | 847,956 | 942,995 | 1,056,254 |
| 1994 | 10,648 | 3,489 | 166,026 | 115,975 | 11,554 | 261,595 | 309,537 | 369,039 | 37,382 | 21,287 | 125,515 | 80,551 | 25,178 | 298,743 | 511,111 | 605,440 | 714,663 | 809,869 | 917,916 | 1,038,560 |
| 1995 | 10,504 | 10,034 | 107,985 | 121,496 | 18,998 | 236,417 | 271,472 | 310,367 | 11,594 | 29,001 | 178,918 | 64,831 | 25,809 | 298,492 | 529,298 | 617,517 | 720,081 | 793,947 | 890,039 | 998,959 |
| 1996 | 30,448 | 5,353 | 80,594 | 138,539 | 11,788 | 237,228 | 268,983 | 304,135 | 14,483 | 25,108 | 183,566 | 49,233 | 34,737 | 227,564 | 464,913 | 543,735 | 635,799 | 725,578 | 814,422 | 911,633 |
| 1997 | 25,893 | 7,329 | 105,152 | 158,309 | 5,406 | 266,473 | 304,401 | 345,076 | 7,486 | 18,346 | 226,390 | 45,732 | 38,490 | 158,702 | 429,392 | 501,583 | 590,932 | 724,416 | 807,238 | 904,865 |
| 1998 | 30,051 | 12,375 | 138,416 | 163,034 | 4,751 | 306,240 | 351,168 | 399,884 | 14,610 | 25,075 | 220,388 | 51,833 | 156,208 | 233,414 | 624,496 | 713,473 | 815,028 | 964,726 | 1,066,027 | 1,176,555 |
| 1999 | 34,351 | 6,574 | 128,344 | 162,781 | 7,038 | 297,429 | 341,153 | 389,292 | 4,858 | 20,757 | 232,224 | 42,319 | 20,053 | 107,933 | 364,933 | 433,104 | 519,208 | 692,273 | 776,449 | 871,865 |
| 2000 | 36,497 | 6,908 | 213,843 | 141,438 | 13,938 | 356,100 | 415,387 | 483,921 | 12,638 | 18,401 | 350,533 | 64,226 | 33,133 | 218,889 | 607,035 | 706,176 | 831,569 | 1,005,927 | 1,124,512 | 1,262,689 |
| 2001 | 16,369 | 6,404 | 185,803 | 198,255 | 9,121 | 357,579 | 419,627 | 491,101 | 10,745 | 16,836 | 256,145 | 57,373 | 31,132 | 220,887 | 523,016 | 602,799 | 692,896 | 920,458 | 1,024,173 | 1,138,376 |
| 2002 | 14,356 | 11,324 | 111,867 | 211,069 | 9,224 | 304,090 | 359,539 | 426,163 | 24,470 | 21,099 | 215,192 | 54,423 | 70,184 | 179,510 | 502,450 | 577,322 | 662,121 | 842,096 | 938,890 | 1,043,251 |
| 2003 | 16,696 | 6,017 | 156,299 | 199,140 | 5,696 | 324,644 | 387,506 | 458,646 | 16,193 | 25,364 | 246,544 | 45,900 | 41,243 | 228,216 | 544,032 | 615,696 | 698,073 | 907,066 | 1,005,212 | 1,113,663 |
| 2004 | 6,592 | 16,532 | 94,026 | 145,987 | 4,921 | 229,347 | 269,644 | 317,994 | 19,247 | 25,178 | 156,351 | 80,943 | 40,839 | 267,169 | 529,533 | 605,419 | 695,016 | 789,341 | 876,664 | 976,395 |
| 2005 | 16,774 | 14,950 | 140,069 | 132,829 | 4,163 | 265,329 | 311,317 | 363,055 | 12,640 | 36,989 | 171,727 | 66,566 | 55,590 | 293,942 | 577,031 | 650,445 | 733,233 | 875,531 | 963,152 | 1,057,897 |
| 2006 | 31,719 | 15,569 | 111,147 | 161,956 | 4,999 | 279,371 | 328,322 | 383,539 | 17,652 | 26,127 | 126,666 | 68,156 | 38,313 | 286,810 | 507,468 | 578,429 | 661,759 | 819,998 | 908,237 | 1,006,826 |
| 2007 | 5,890 | 11,690 | 61,869 | 122,760 | 2,385 | 173,407 | 206,199 | 246,987 | 13,869 | 30,651 | 251,219 | 66,028 | 75,189 | 285,228 | 631,887 | 743,134 | 947,044 | 833,487 | 952,338 | 1,155,632 |
| 2008 | 6,046 | 11,078 | 87,973 | 93,257 | 3,589 | 173,146 | 203,737 | 237,234 | 13,609 | 37,449 | 245,225 | 64,882 | 43,695 | 251,479 | 567,390 | 681,428 | 879,013 | 767,933 | 887,062 | 1,083,689 |
| 2009 | 12,349 | 18,475 | 71,264 | 100,857 | 3,816 | 178,703 | 209,016 | 245,126 | 4,879 | 41,304 | 203,631 | 40,616 | 34,896 | 217,663 | 469,441 | 562,972 | 721,921 | 673,452 | 773,506 | 935,463 |
| 2010 | 11,553 | 14,827 | 115,899 | 92,556 | 6,433 | 208,546 | 243,809 | 283,732 | 16,508 | 42,876 | 257,911 | 80,816 | 32,204 | 391,691 | 708,336 | 859,613 | 1,088,911 | 948,261 | 1,104,333 | 1,334,069 |
| 2011 | 14,127 | 12,605 | 80,386 | 102,550 | 4,024 | 185,699 | 215,400 | 248,904 | 11,640 | 30,363 | 227,170 | 44,952 | 28,362 | 206,916 | 467,408 | 568,936 | 763,230 | 678,281 | 786,470 | 981,319 |
| 2012 | 30,689 | 4,824 | 90,171 | 109,583 | 5,732 | 209,805 | 243,422 | 280,030 | 10,090 | 18,539 | 230,049 | 29,160 | 37,810 | 302,500 | 531,785 | 658,798 | 859,867 | 770,531 | 903,561 | 1,109,216 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 yr Av . | 15,244 | 12,657 | 100,910 | 126,148 | 4,576 | 222,800 | 261,837 | 306,525 | 13,633 | 31,484 | 211,649 | 58,802 | 42,814 | 273,161 | 553,431 | 652,487 | 804,807 | 806,388 | 916,053 | 1,075,417 |

Table 3.3.5.6 Estimated number of MSW SPAWNERS by NEAC country or region and year

|  | Northern Europe |  |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  |  |  | NEAC Area |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Finland | Iceland | Norway | Russia | Sweden | Total |  |  | France | Iceland | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  |  | Total |  |  |
|  |  | N\&E |  |  |  | 5,0\% | 50,0\% | 95,0\% |  | N\&E |  |  |  |  | 5,0\% | 50,0\% | 95,0\% | 5,0\% | 50,0\% | 95,0\% |
| 1971 | 10789 | 2905 |  | NA | 444 |  |  |  | 6765 | 7287 | 82117 | 52014 | 10991 | 306687 | 386943 | 474087 | 577548 |  |  |  |
| 1972 | 16831 | 4530 |  | 59083 | 313 |  |  |  | 13415 | 11212 | 88679 | 92879 | 9594 | 388784 | 499583 | 615322 | 752434 |  |  |  |
| 1973 | 20058 | 4219 |  | 66122 | 1083 |  |  |  | 8285 | 10193 | 96422 | 71563 | 8380 | 436521 | 518379 | 640325 | 789200 |  |  |  |
| 1974 | 29906 | 4011 |  | 98388 | 702 |  |  |  | 3900 | 8786 | 108329 | 52866 | 9131 | 284175 | 381865 | 476317 | 588192 |  |  |  |
| 1975 | 33286 | 4454 |  | 86775 | 167 |  |  |  | 7668 | 9301 | 121625 | 72189 | 7507 | 311047 | 428952 | 541550 | 671436 |  |  |  |
| 1976 | 27314 | 3628 |  | 86244 | 511 |  |  |  | 5606 | 8023 | 83999 | 37701 | 5233 | 224978 | 301606 | 372708 | 452379 |  |  |  |
| 1977 | 16769 | 5063 |  | 71674 | 383 |  |  |  | 4364 | 7868 | 73307 | 48068 | 5153 | 208770 | 284486 | 355052 | 438938 |  |  |  |
| 1978 | 10650 | 6524 |  | 50573 | 294 |  |  |  | 4452 | 10168 | 62695 | 40901 | 6709 | 287365 | 337764 | 419889 | 517009 |  |  |  |
| 1979 | 13885 | 4311 |  | 44365 | 848 |  |  |  | 5065 | 6490 | 57519 | 20543 | 4704 | 202385 | 238626 | 301575 | 378815 |  |  |  |
| 1980 | 14531 | 6011 |  | 47703 | 1489 |  |  |  | 10685 | 9101 | 62906 | 66983 | 5959 | 242791 | 329441 | 406651 | 503557 |  |  |  |
| 1981 | 16017 | 2116 |  | 66054 | 430 |  |  |  | 7540 | 6090 | 46623 | 94265 | 4664 | 255922 | 339448 | 424557 | 524969 |  |  |  |
| 1982 | 21158 | 2423 |  | 40652 | 1558 |  |  |  | 4735 | 4299 | 32709 | 36696 | 6755 | 237138 | 268959 | 325779 | 403186 |  |  |  |
| 1983 | 22477 | 1847 | 101267 | 49141 | 1068 | 142935 | 178501 | 218920 | 5034 | 7176 | 111954 | 41965 | 9489 | 243457 | 330428 | 435297 | 631108 | 502797 | 616015 | 814188 |
| 1984 | 21418 | 2384 | 103995 | 62180 | 1492 | 157626 | 193580 | 233864 | 8111 | 6050 | 43138 | 33432 | 3719 | 223924 | 269433 | 322333 | 385269 | 452863 | 517354 | 591901 |
| 1985 | 16822 | 1542 | 95793 | 51199 | 624 | 135591 | 167309 | 202557 | 6179 | 4404 | 53412 | 49212 | 4830 | 296310 | 353413 | 417893 | 496938 | 512509 | 586747 | 672686 |
| 1986 | 14660 | 4151 | 115156 | 52222 | 604 | 149543 | 188827 | 232783 | 6340 | 3672 | 51048 | 67371 | 5434 | 379843 | 434094 | 520415 | 626439 | 614212 | 710003 | 822602 |
| 1987 | 18280 | 4325 | 89795 | 53178 | 1814 | 137749 | 169718 | 206292 | 3331 | 3263 | 79647 | 55123 | 2993 | 244470 | 331710 | 394079 | 466714 | 493340 | 565684 | 645985 |
| 1988 | 12060 | 2802 | 72290 | 44822 | 1767 | 111116 | 135245 | 163017 | 9279 | 3741 | 53160 | 71967 | 10046 | 442765 | 510793 | 596951 | 701724 | 641454 | 733191 | 840838 |
| 1989 | 10956 | 2379 | 77794 | 50856 | 4880 | 126808 | 148345 | 172604 | 4211 | 3325 | 40876 | 57102 | 4992 | 389772 | 431811 | 505326 | 593292 | 576271 | 654129 | 745363 |
| 1990 | 13726 | 2501 | 91137 | 48257 | 3719 | 135760 | 160762 | 189884 | 4361 | 3276 | 14906 | 71161 | 7024 | 311274 | 350606 | 417473 | 494108 | 506892 | 579565 | 659623 |
| 1991 | 16354 | 1728 | 76485 | 60517 | 4279 | 137984 | 161194 | 186992 | 3920 | 3275 | 41168 | 31705 | 3315 | 249857 | 290029 | 336299 | 393170 | 445549 | 497985 | 559946 |
| 1992 | 17623 | 2593 | 84500 | 58618 | 5474 | 146406 | 170315 | 197506 | 4902 | 3700 | 20871 | 24225 | 8915 | 344188 | 351251 | 409438 | 482756 | 516693 | 581030 | 657573 |
| 1993 | 20397 | 2919 | 78173 | 55922 | 7459 | 142410 | 166793 | 193204 | 2323 | 1812 | 24442 | 27644 | 27666 | 273997 | 310668 | 363966 | 428345 | 471694 | 531308 | 600814 |
| 1994 | 16944 | 2493 | 76900 | 65296 | 5537 | 144498 | 168525 | 194497 | 5343 | 2937 | 40170 | 39199 | 6624 | 335903 | 371340 | 434087 | 507993 | 535203 | 603497 | 681633 |
| 1995 | 10501 | 1704 | 83514 | 64449 | 4422 | 141214 | 165903 | 193764 | 2546 | 3325 | 37969 | 40631 | 5440 | 306798 | 342269 | 400087 | 471517 | 502354 | 566694 | 643183 |
| 1996 | 11273 | 2272 | 83038 | 63445 | 5697 | 142312 | 167144 | 193821 | 4511 | 2144 | 19580 | 42577 | 6776 | 241212 | 272680 | 321091 | 379022 | 433652 | 489337 | 553529 |
| 1997 | 16340 | 1268 | 57823 | 52666 | 3662 | 113571 | 133438 | 155554 | 2338 | 2391 | 39204 | 27400 | 8421 | 164793 | 209920 | 250983 | 299105 | 338740 | 384857 | 437621 |
| 1998 | 13947 | 1855 | 69775 | 41824 | 2707 | 111049 | 131691 | 154079 | 1961 | 1493 | 12544 | 18005 | 13586 | 181317 | 199195 | 231575 | 270794 | 324439 | 364313 | 408711 |
| 1999 | 11837 | 2484 | 72209 | 54654 | 2331 | 121751 | 144001 | 169040 | 4270 | 3109 | 33726 | 38337 | 5416 | 133942 | 184006 | 229121 | 284207 | 322482 | 373839 | 434459 |
| 2000 | 26395 | 1504 | 102507 | 58877 | 5102 | 167025 | 195914 | 228237 | 2971 | 898 | 44010 | 40986 | 7186 | 175979 | 237346 | 278054 | 327296 | 423374 | 474923 | 532077 |
| 2001 | 37870 | 1813 | 122415 | 89635 | 6122 | 222967 | 259929 | 300592 | 3499 | 1517 | 36725 | 44608 | 5474 | 167809 | 223618 | 266311 | 317515 | 468393 | 527920 | 591455 |
| 2002 | 30033 | 1806 | 107133 | 74466 | 4472 | 188004 | 219715 | 256655 | 3232 | 1745 | 47471 | 40197 | 5284 | 139332 | 206609 | 244362 | 289217 | 414658 | 465034 | 522180 |
| 2003 | 21513 | 2231 | 95610 | 63369 | 5085 | 162260 | 189846 | 220017 | 4638 | 2557 | 54415 | 53637 | 3095 | 184750 | 262860 | 310431 | 368696 | 444916 | 501401 | 565419 |
| 2004 | 10302 | 2095 | 87575 | 48196 | 3728 | 129568 | 153219 | 180534 | 8624 | 2129 | 24727 | 46037 | 3080 | 238605 | 280003 | 330089 | 389422 | 428046 | 484123 | 548343 |
| 2005 | 7956 | 2667 | 79267 | 36445 | 2834 | 109381 | 129757 | 152520 | 5309 | 1987 | 37627 | 49465 | 4194 | 188774 | 250206 | 293259 | 346535 | 376312 | 423597 | 480674 |
| 2006 | 13774 | 3042 | 101037 | 46489 | 2816 | 142595 | 168570 | 197597 | 5430 | 1651 | 25234 | 45516 | 2906 | 198926 | 242163 | 288302 | 342602 | 403373 | 457328 | 519600 |
| 2007 | 19905 | 3376 | 84031 | 39782 | 3925 | 129595 | 152053 | 176533 | 5077 | 986 | 14158 | 44619 | 4800 | 193231 | 226881 | 269309 | 321396 | 372985 | 421662 | 479222 |
| 2008 | 19021 | 3749 | 125852 | 47364 | 6588 | 173333 | 204001 | 240190 | 5629 | 1426 | 21223 | 49347 | 2777 | 219208 | 257869 | 306829 | 368205 | 452041 | 512771 | 581566 |
| 2009 | 8762 | 3533 | 100318 | 69891 | 5919 | 162052 | 190857 | 223951 | 2945 | 1902 | 23503 | 37408 | 3869 | 188547 | 221592 | 264313 | 315597 | 403074 | 455952 | 516989 |
| 2010 | 13917 | 4849 | 122947 | 60858 | 7229 | 180809 | 211566 | 245297 | 2488 | 3729 | 15111 | 55772 | 3556 | 244716 | 275224 | 333757 | 407867 | 478898 | 546269 | 626921 |
| 2011 | 10922 | 5760 | 178797 | 72588 | 6999 | 234896 | 276784 | 324242 | 6468 | 2064 | 17379 | 78825 | 9483 | 274152 | 327901 | 400463 | 495720 | 592713 | 678972 | 784021 |
| 2012 | 14165 | 2821 | 156484 | 64324 | 11069 | 214910 | 251039 | 291568 | 5046 | 1907 | 17645 | 65175 | 12482 | 234750 | 283044 | 347497 | 428407 | 524247 | 600008 | 689126 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 3.3.7.1. Estimated survival of wild smolts (\%) to return to homewaters (prior to coastal fisheries) for various monitored rivers in the NE Atlantic Area.

| Smolt | Iceland ${ }^{1}$ |  |  | Norway ${ }^{2}$ |  |  |  | Ireland |  | $\begin{gathered} \text { B'shoole } \\ \hline \text { 1SW } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ellidaar | R.Vesturdalsa ${ }^{4}$ |  | R. Halselva |  | R. Imsa |  | R. Corrib |  |  |
| year | 1sw | 1sw | 2SW | 1SW | 2SW | 1SW | 2SW | 1SW | 2SW |  |
| 1975 | 20.8 |  |  |  |  |  |  |  |  |  |
| 1980 |  |  |  |  |  |  |  | 17.9 | 1.1 | 3.1 |
| 1981 |  |  |  |  |  | 17.3 | 4.0 | 9.2 | 3.8 | 5.4 |
| 1982 |  |  |  |  |  | 5.3 | 1.2 | 20.9 | 3.3 | 5.8 |
| 1983 |  |  |  |  |  | 13.5 | 1.3 | 10.0 | 1.8 | 3.4 |
| 1984 |  |  |  |  |  | 12.1 | 1.8 | 26.2 | 2.0 | 7.8 |
| 1985 | 9.4 |  |  |  |  | 10.2 | 2.1 | 18.9 | 1.8 | 7.9 |
| 1986 |  |  |  |  |  | 3.8 | 4.2 |  |  | 8.7 |
| 1987 |  |  |  | 2.0 | 0.3 | 17.3 | 5.6 | 16.6 | 0.7 | 12.0 |
| 1988 | 12.7 |  |  | 5.8 | 0.7 | 13.3 | 1.1 | 14.6 | 0.7 | 10.1 |
| 1989 | 8.1 |  |  | 2.1 | 1.0 | 8.7 | 2.2 | 6.7 | 0.7 | 3.5 |
| 1990 | 5.4 |  |  | 3.9 | 1.6 | 3.0 | 1.3 | 5.0 | 0.6 | 9.2 |
| 1991 | 8.8 |  |  | 2.1 | 0.3 | 8.7 | 1.2 | 7.3 | 1.3 | 9.5 |
| 1992 | 9.6 |  |  | 2.1 | 0.4 | 6.7 | 0.9 | 7.3 |  | 7.6 |
| 1993 | 9.8 |  |  | 2.1 | 0.0 | 15.6 |  | 10.8 | 0.1 | 9.5 |
| 1994 | 9.0 |  |  | 0.6 | 0.4 |  |  | 9.8 | 1.4 | 9.4 |
| 1995 | 9.4 |  | 1.5 | 0.9 | 0.0 | 1.8 | 1.5 | 8.4 | 0.1 | 6.8 |
| 1996 | 4.6 | 2.5 | 0.4 | 2.8 | 0.6 | 3.5 | 0.9 | 6.3 | 1.2 | 9.2 |
| 1997 | 5.3 | 1.0 | 1.5 | 0.8 | 0.0 | 1.7 | 0.3 | 12.7 | 0.8 | 8.2 |
| 1998 | 5.3 | 1.5 | 1.0 | 1.5 | 0.6 | 7.2 | 1.0 | 5.5 | 1.1 | 5.3 |
| 1999 | 7.7 | 1.3 | 1.2 | 1.3 | 0.0 | 4.2 | 2.2 | 6.4 | 0.9 | 8.1 |
| 2000 | 6.3 | 1.1 | 0.7 | 0.4 | 1.1 | 12.5 | 1.7 | 9.4 | 0.0 | 9.0 |
| 2001 | 5.1 | 3.4 | 1.3 | 2.5 | 2.5 | 3.6 | 2.2 | 7.2 | 1.1 | 7.6 |
| 2002 | 4.4 | 1.1 | 2.3 | 0.8 | 0.6 | 5.5 | 0.9 | 6.0 | 0.5 | 6.5 |
| 2003 | 9.1 | 5.5 | 0.6 | 4.9 | 1.6 | 3.5 | 0.7 | 8.3 | 2.1 | 8.3 |
| 2004 | 7.7 | 5.7 | 0.6 | 3.5 | 1.2 | 5.9 | 1.4 | 6.3 | 0.8 | 5.8 |
| 2005 | 6.4 | 2.5 | 0.9 | 3.0 | 1.0 | 3.7 | 1.8 |  | 0.0 | 5.3 |
| 2006 | 7.1 | 1.8 | 1.0 | 0.8 | 0.8 | 0.8 | 5.8 | 1.2 | 0.9 | 13.0 |
| 2007 | 19.3 | 0.9 | 0.3 | 0.3 | 0.0 | 0.8 | 0.6 | 0.9 | 0.0 | 8.4 |
| 2008 | 14.9 | 2.6 | 1.1 | 0.2 | 0.2 | 1.1 | 2.3 | 1.7 | 1.03 | 8.22 |
| 2009 | 14.2 | 1.3 | 1.6 | 1.1 | 0.6 | 2.4 | 3.1 | 6 |  | 8.85 |
| 2010 | 8.6 | 2.0 | 1.1 | 0.4 | 0.4 | 1.7 | 1.3 |  |  | 7.49 |
| 2011 | 6.1 | 1.3 |  | 0 |  | 4.5 |  |  |  | 10.81 |
| Mean |  |  |  |  |  |  |  |  |  |  |
| (5-year) | 12.8 | 1.7 | 1.0 | 0.6 | 0.5 | 1.4 | 2.7 | 2.5 | 0.5 | 8.7 |
| (10-year) | 9.7 | 2.7 | 1.0 | 1.8 | 1.0 | 2.9 | 2.1 | 4.7 | 0.7 | 8.1 |
|  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{1}$ Microtags. |  |  |  |  |  | ${ }^{5}$ From 0+ stage in autumn. |  |  |  |  |
| ${ }^{2}$ Carlin tags, not corrected for tagging mortality. |  |  |  |  |  | ${ }^{6}$ Incomplete returns. |  |  |  |  |
| ${ }^{3}$ Microtags, corrected for tagging mortality. |  |  |  |  |  | ${ }^{7}$ Assumes $30 \%$ exploitation in trap fishery. |  |  |  |  |
| ${ }^{4}$ Assumes 50\% exploitation in rod fishery. |  |  |  |  |  | ${ }^{8}$ France data based on retruns to freshwater |  |  |  |  |

Table 3.3.7.1. Cont'd. Estimated survival of wild smolts (\%) to return to homewaters (prior to coastal fisheries) for various monitored rivers in the NE Atlantic Area.


Table 3.3.7.2. Estimated survival of hatchery smolts (\%) to return to homewaters (prior to coastal fisheries) for monitored rivers and experimental facilities in the NE Atlantic Area.

|  | Iceland ${ }^{1}$ <br> R. Ranga |  | Norway ${ }^{2}$ |  |  |  |  |  | Sweden ${ }^{2}$ <br> R. Lagan |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Smolt year |  |  | R. Halselva |  | R. Imsa |  | R. Drammen |  |  |  |
|  | 1SW | 2SW | 1SW | 2SW | 1SW | 2SW | 1SW | 2SW | 1SW | 2SW |
| 1981 |  |  |  |  | 10.1 | 1.3 |  |  |  |  |
| 1982 |  |  |  |  | 4.2 | 0.6 |  |  |  |  |
| 1983 |  |  |  |  | 1.6 | 0.1 |  |  |  |  |
| 1984 |  |  |  |  | 3.8 | 0.4 | 3.5 | 3.0 | 11.8 | 1.1 |
| 1985 |  |  |  |  | 5.8 | 1.3 | 3.4 | 1.9 | 11.8 | 0.9 |
| 1986 |  |  |  |  | 4.7 | 0.8 | 6.1 | 2.2 | 7.9 | 2.5 |
| 1987 |  |  | 1.5 | 0.4 | 9.8 | 1.0 | 1.7 | 0.7 | 8.4 | 2.4 |
| 1988 |  |  | 1.2 | 0.1 | 9.5 | 0.7 | 0.5 | 0.3 | 4.3 | 0.6 |
| 1989 | 1.6 | 0.1 | 1.9 | 0.5 | 3.0 | 0.9 | 1.9 | 1.3 | 5.0 | 1.3 |
| 1990 | 0.8 | 0.2 | 2.1 | 0.3 | 2.8 | 1.5 | 0.3 | 0.4 | 5.2 | 3.1 |
| 1991 | 0.0 | 0.0 | 0.6 | 0.0 | 3.2 | 0.7 | 0.1 | 0.1 | 3.6 | 1.1 |
| 1992 | 0.4 | 0.1 | 0.5 | 0.0 | 3.8 | 0.7 | 0.4 | 0.6 | 1.5 | 0.4 |
| 1993 | 0.7 | 0.1 |  |  | 6.5 | 0.5 | 3.0 | 1.0 | 2.6 | 0.9 |
| 1994 | 1.2 | 0.2 |  |  | 6.2 | 0.6 | 1.2 | 0.9 | 4.0 | 1.2 |
| 1995 | 1.1 | 0.1 |  |  | 0.4 | 0.0 | 0.7 | 0.3 | 3.9 | 0.6 |
| 1996 | 0.2 | 0.0 | 1.2 | 0.2 | 2.1 | 0.2 | 0.3 | 0.2 | 3.5 | 0.5 |
| 1997 | 0.3 | 0.1 | 0.6 | 0.0 | 1.0 | 0.0 | 0.5 | 0.2 | 0.6 | 0.5 |
| 1998 | 0.5 | 0.0 | 0.5 | 0.5 | 2.4 | 0.1 | 1.9 | 0.7 | 1.6 | 0.9 |
| 1999 | 0.4 | 0.0 | 2.3 | 0.2 | 12.0 | 1.1 | 1.9 | 1.6 | 2.1 |  |
| 2000 | 0.9 | 0.1 | 1.0 | 0.7 | 8.4 | 0.1 | 1.1 | 0.6 |  |  |
| 2001 | 0.4 | 0.1 | 1.9 | 0.6 | 3.3 | 0.3 | 2.5 | 1.1 |  |  |
| 2002 | 0.4 |  | 1.4 | 0.0 | 4.5 | 0.8 | 1.2 | 0.8 |  |  |
| 2003 | 0.2 |  | 0.5 | 0.3 | 2.6 | 0.7 | 0.3 | 0.6 |  |  |
| 2004 | 0.6 |  | 0.2 | 0.1 | 3.6 | 0.7 | 0.4 | 0.4 |  |  |
| 2005 | 1.0 |  | 1.2 | 0.2 | 2.8 | 1.2 | 0.3 | 0.7 |  |  |
| 2006 | 1.0 |  | 0.0 | 0.1 | 1.0 | 1.8 | 0.1 | 0.6 |  |  |
| 2007 | 1.9 |  | 0.3 | 0.0 | 0.6 | 0.7 | 0.2 | 0.1 |  |  |
| 2008 | 2.4 |  | 0.1 | 0 | 1.8 | 2.2 | 0.1 | 0.3 |  |  |
| 2009 | 0.0 |  | 0 | 0 | 1.3 | 3.3 | 0 | 0 |  |  |
| 2010 | 0.5 |  | 0.8 |  | 3.7 | 1.8 | 0 |  |  |  |
| 2011 | 0.5 |  |  |  | 2.3 |  |  |  |  |  |
| Mean |  |  |  |  |  |  |  |  |  |  |
| (5-year) | 1.2 |  | 0.2 |  | 1.5 | 1.8 | 0.1 | 0.3 |  |  |
| (10-year) | 0.8 |  | 0.6 |  | 2.5 | 1.2 | 0.5 | 0.5 |  |  |

Table 3.3.7.2. Cont'd. Estimated survival of hatchery smolts (\%) to return to homewaters (prior to coastal fisheries) for monitored rivers and experimental facilities in the NE Atlantic Area.

|  | Ireland |  |  |  |  |  |  |  |  |  | UK (N. Ireland) ${ }^{3}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Smolt year | R. Shannon | R. Screebe | R. Burrishoole $^{1}$ | R. Delphi/ R. Burrishoole ${ }^{4}$ | R. Delphi | $\begin{gathered} \mathrm{R} . \\ \text { Bunowen } \end{gathered}$ | R. Lee | R. Corrib Cong. ${ }^{2}$ | R. Corrib Galway ${ }^{2}$ | R. Erne | $\begin{aligned} & \text { R. Bush } \\ & 1+\text { smolts } \end{aligned}$ | R. Bush 2+ smolts |
| 1980 | 8.6 |  | 5.6 |  |  |  | 8.3 | 0.9 |  |  |  |  |
| 1981 | 2.8 |  | 8.1 |  |  |  | 2.0 | 1.5 |  |  |  |  |
| 1982 | 4.0 |  | 11.0 |  |  |  | 16.3 | 2.7 | 0.4 |  |  |  |
| 1983 | 3.9 |  | 4.6 |  |  |  |  | 2.8 |  |  | 1.9 | 8.1 |
| 1984 | 5.0 | 10.4 | 27.1 |  |  |  | 2.3 | 5.2 |  | 9.4 | 13.3 |  |
| 1985 | 17.8 | 12.3 | 31.1 |  |  |  | 15.7 | 1.4 |  | 8.2 | 15.4 | 17.5 |
| 1986 | 2.1 | 0.4 | 9.4 |  |  |  | 16.4 |  |  | 10.8 | 2.0 | 9.7 |
| 1987 | 4.7 | 8.4 | 14.1 |  |  |  | 8.8 |  |  | 7.0 | 6.5 | 19.4 |
| 1988 | 4.9 | 9.2 | 17.2 |  |  |  | 5.5 | 4.5 |  | 2.9 | 4.9 | 6.0 |
| 1989 | 5.0 | 1.8 | 10.5 |  |  |  | 1.7 | 6.0 |  | 1.2 | 8.1 | 23.2 |
| 1990 | 1.3 |  | 11.4 |  | 0.2 |  | 2.5 | 0.2 | 16.1 | 2.6 | 5.6 | 5.6 |
| 1991 | 4.2 | 0.3 | 13.6 | 10.8 | 6.2 |  | 0.8 | 4.9 | 4.1 | 1.3 | 5.4 | 8.8 |
| 1992 | 4.4 | 1.3 | 7.4 | 10.0 | 1.7 | 4.2 |  | 0.9 | 13.2 |  | 6.0 | 7.8 |
| 1993 | 2.9 | 3.4 | 12.0 | 14.3 | 6.5 | 5.4 |  | 1.0 | 14.5 |  | 1.1 | 5.8 |
| 1994 | 5.2 | 1.9 | 14.3 | 3.9 | 2.7 | 10.8 |  |  | 7.7 |  | 1.6 |  |
| 1995 | 3.6 | 4.1 | 6.6 | 3.4 | 1.7 | 3.5 |  | 2.4 | 2.2 |  | 3.1 | 2.4 |
| 1996 | 2.9 | 1.8 | 5.3 | 10.6 | 6.7 | 3.4 |  |  |  |  | 2.0 | 2.3 |
| 1997 | 6.0 | 0.4 | 13.3 | 17.3 | 5.6 | 5.3 | 7.0 |  | 4.8 | 7.7 | - | 4.1 |
| 1998 | 3.1 | 1.3 | 4.9 | 7.2 | 3.1 | 2.9 | 4.9 | 3.3 | 2.3 | 2.6 | 2.3 | 4.5 |
| 1999 | 1.0 | 2.8 | 8.2 | 19.9 | 8.2 | 2.0 |  |  | 4.0 | 3.3 | 2.7 | 5.8 |
| 2000 | 1.2 | 3.8 | 11.8 | 19.5 | 13.2 | 5.4 | 3.55 | 6.7 |  | 4.0 | 2.8 | 4.4 |
| 2001 | 2.0 | 2.5 | 9.7 | 17.2 | 7.4 | 3.2 | 1.95 | 3.4 |  | 6.0 | 1.1 | 2.2 |
| 2002 | 1.0 | 4.1 | 9.2 | 12.6 | 4.9 | 2.0 | 1.93 |  | 5.3 | 1.9 | 0.7 | 3.1 |
| 2003 | 1.2 |  | 6.0 | 3.7 | 1.5 | 1.6 | 4.31 |  |  | 1.0 | 2.5 | 1.9 |
| 2004 | 0.4 | 1.8 | 9.4 | 7.6 | 2.3 | 1.8 | 2.23 |  |  | 3.1 | 0.7 | 1.9 |
| 2005 | 0.6 | 3.4 | 4.9 | 11.0 |  | 1.0 | 0.96 |  |  | 0.9 | 1.8 | 1.7 |
| 2006 | 0.3 | 1.3 | 5.2 | 3.7 | 1.5 | 0.02 | 0.19 | 0.4 | 2.9 | 0.9 | 2.0 | 3.8 |
| 2007 | 0.5 | 0.8 | 7.1 | 0.0 | 3.6 |  |  |  | 3.6 | 0.7 |  |  |
| 2008 |  | 0.2 | 1.3 | 0.0 | 1.4 |  | 0.05 |  |  |  |  |  |
| 2009 | 0.3 | 0.2 | 2.3 | 0.0 | 1.5 |  | 0.07 |  |  | 1.1 |  |  |
| 2010 | 0.4 | 0.4 | 2.8 |  | 3.6 |  | 0.13 |  | 2.0 | 0.9 |  |  |
| 2011 |  |  | 4.7 |  |  |  |  |  |  |  | 0.8 | 1.86 |
| Mean |  |  |  |  |  |  |  |  |  |  |  |  |
| (5-year) | 0.4 | 0.6 | 3.8 | 0.9 | 2.3 | 0.0 | 0.1 | 0.4 | 2.8 | 0.9 | 2.0 | 3.8 |
| (10-year) | 0.7 | 1.6 | 5.8 | 6.2 | 3.1 | 1.6 | 1.3 | 1.9 | 3.4 | 1.8 | 1.5 | 2.4 |
| ${ }^{1}$ Return rates to rod fishery with con |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{2}$ Different release sites |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{3}$ Microtagged. |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{4}$ Delphi fish released at Burrishoole |  |  |  |  |  |  |  |  |  |  |  |  |

Table 3.3.6.1. Compliance with river-specific conservation limits for individual river stocks, after homewater fisheries, by jurisdiction in the NEAC area in 2012 (except Norway where data are for 2010). Data for France are for MSW fish only.

| Country | No. RIVERS | No. WITH CL | No. ASESSED FOR <br> COMPLIANCE | No. <br> ATTAINING CL |
| :--- | :---: | :---: | :---: | :---: | :---: |

Table 3.4.5.1. Catch in weight $(t)$ and numbers, mean weights and mean age of catch in the 1983/1984 to 1995/1996 fishing seasons.

|  | Season | Catch (t) | Catch <br> (number) | Mean wt <br> $(\mathbf{k g})$ | Mean sea <br> age |
| :--- | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |
| Commercial | $1983 / 84$ | 651 | 124509 | 5.23 | 2.07 |
| fishery | $1984 / 85$ | 598 | 135777 | 4.40 | 2.07 |
|  | $1985 / 86$ | 545 | 154554 | 3.53 | 2.02 |
|  | $1986 / 87$ | 539 | 140304 | 3.84 | 2.05 |
|  | $1987 / 88$ | 208 | 65011 | 3.20 | 1.96 |
|  | $1988 / 89$ | 309 | 93496 | 3.30 | 2.04 |
|  | $1989 / 90$ | 364 | 111515 | 3.26 | 2.04 |
|  | $1990 / 91$ | 202 | 57441 | 3.52 | 2.07 |
| Research | $1991 / 92$ | 31 | 8464 | 3.66 | 2.09 |
| fishery | $1992 / 93$ | 22 | 5415 | 4.06 | 2.14 |
|  | $1993 / 94$ | 7 | 2072 | 3.38 | 2.03 |
|  | $1994 / 95$ | 6 | 1963 | 3.06 | 1.98 |
|  | $1995 / 96$ | 1 | 282 | 3.55 |  |

Table 3.4.5.2. Catch in numbers and percentages by sea age and mean age in the Faroes salmon fishery in the 1983/1984 to 1994/1995 seasons.

| Fishery | Season | 1SW | 2SW | 3SW | MSW | \%1SW | \%2SW | \%3SW | Mean Age |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Commercial | $1983 / 84$ | 5142 | 135718 | 16401 | 152178 | $3.3 \%$ | $86.3 \%$ | $10.4 \%$ | 2.07 |
| fishery | $1984 / 85$ | 381 | 138375 | 11358 | 149733 | $0.3 \%$ | $92.2 \%$ | $7.6 \%$ | 2.07 |
|  | $1985 / 86$ | 2021 | 169461 | 5671 | 175219 | $1.1 \%$ | $95.7 \%$ | $3.2 \%$ | 2.02 |
|  | $1986 / 87$ | 71 | 124628 | 6621 | 131324 | $0.1 \%$ | $94.9 \%$ | $5.0 \%$ | 2.05 |
|  | $1987 / 88$ | 5833 | 55726 | 3450 | 59176 | $9.0 \%$ | $85.7 \%$ | $5.3 \%$ | 1.96 |
|  | $1988 / 89$ | 1351 | 110717 | 5728 | 116445 | $1.1 \%$ | $94.0 \%$ | $4.9 \%$ | 2.04 |
|  | $1989 / 90$ | 2155 | 102800 | 6473 | 109273 | $1.9 \%$ | $92.3 \%$ | $5.8 \%$ | 2.04 |
|  | $1990 / 91$ | 632 | 52419 | 4390 | 56809 | $1.1 \%$ | $91.3 \%$ | $7.6 \%$ | 2.07 |
| Research | $1991 / 92$ | 248 | 4686 | 743 | 5429 | $4.4 \%$ | $82.5 \%$ | $13.1 \%$ | 2.09 |
| fishery | $1992 / 93$ | 521 | 2646 | 1120 | 3766 | $12.2 \%$ | $61.7 \%$ | $26.1 \%$ | 2.14 |
|  | $1993 / 94$ | 320 | 1288 | 376 | 1664 | $16.1 \%$ | $64.9 \%$ | $19.0 \%$ | 2.03 |
|  | $1994 / 95$ | 206 | 1585 | 166 | 1751 | $10.5 \%$ | $81.0 \%$ | $8.5 \%$ | 1.98 |
|  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  | 2.04 |

1991/92 to 1994/95 include discards and exclude reared fish.

Table 3.4.5.3. Estimates of discard rates in the Faroes fishery in the 1982/1983 to 1994/1995 fishing seasons.

|  | Season | Number of samples | Number sampled | Number \& 60 cm TL | Disard rate <br> (\%) | Range \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commeróal | 1982/83 | 7 | 6820 | 472 | 6.9\% | 0.0\% - 10.4\% |
| fishery | 1983/84 | 5 | 4467 | 176 | 3.9\% | - |
|  | 1984/85 | 12 | 9545 | 1289 | 13.5\% | 3.0\% - $32.0 \%$ |
|  | 1985/86 | 7 | 14654 | 286 | $2.0 \%$ | 0.6\% - 13.8\% |
|  | 1986/87 | 13 | 39758 | 2849 | 7.2\% | 0.0\% - 71.3\% |
|  | 1987/88 | 2 | 1499 | 235 | 15.7\% | - |
|  | 1988/89 | 9 | 17235 | 1804 | 10.5\% | 0.4\% - $31.9 \%$ |
|  | 1989/90 | 5 | 16375 | 1533 | 9.47\% | 3.6\% - 18.5\% |
|  | 1990/91 | 3 | 4515 | 681 | 14.8\% | 9.9\% - 17.5\% |
| Research | 1991/92 | $\epsilon$ | 9350 | 825 | 8.8\% | 2.4\% - 15.9\% |
| fishery | 1992/93 | 3 | 9099 | 853 | 9.47\% | 5.1\% - $32.3 \%$ |
|  | 1993/94 | 4 | 3035 | 436 | 14.4\% | 15\% - $48.6 \%$ |
|  | 1994/95 | 5 | 4187 | 634 | 15.1\% | 5.0\% - 39.7\% |

Table 3.4.5.4. Percentages of farm escapees observed in catch samples taken in the Faroes fishery (1981/1982 to 1995/1996) and the Norwegian coastal fisheries (1989 to 2008).

| Year | Norway coastal fisheries | Season | Faroes fishery <br> (ICES 1996) |
| :---: | :---: | :---: | :---: |
| 1981 |  | 1981/82 | 2 |
| 1982 |  | 1982/83 | 2 |
| 1983 |  | 1983/84 | 1 |
| 1984 |  | 1984/85 | 4 |
| 1985 |  | 1985/86 | 7 |
| 1986 |  | 1986/87 | 4 |
| 1987 |  | 1987/88 | 1 |
| 1988 |  | 1988/89 | 8 |
| 1989 | 45 | 1989/90 | 17 |
| 1990 | 48 | 1990/91 | 43 |
| 1991 | 49 | 1991/92 | 42 |
| 1992 | 44 | 1992/93 | 37 |
| 1993 | 47 | 1993/94 | 27 |
| 1994 | 34 | 1995/95 | 17 |
| 1995 | 42 | 1995/96 | 19 |
| 1996 | 54 |  |  |
| 1997 | 47 |  |  |
| 1998 | 45 |  |  |
| 1999 | 35 |  |  |
| 2000 | 31 |  |  |
| 2001 | 27 |  |  |
| 2002 | 33 |  |  |
| 2003 | 21 |  |  |
| 2004 | 27 |  |  |
| 2005 | 23 |  |  |
| 2006 | 33 |  |  |
| 2007 | 32 |  |  |
| 2008 | 26 |  |  |

Table 3.4.5.5. Additional parameter values used in the example catch advice for the Faroes fishery.

# Minimum TAC option (t) <br> 0 <br> Maximum TAC option (t) <br> ..... 200 <br> TAC steps (t) <br> ..... 20 

Faroes share allocation ..... 0.084
TAC in current year ( $\mathbf{t}$ ) ..... 0
Proportion of 1SW salmon not maturing ..... 0.22
0.8
Monthly rate of natural mortality ..... 0.03

Table 3.5.3.1. Probabilities that the forecast PFA for 1SW maturing and 1SW non-maturing fish will be greater than the age specific Spawner Escapement Reserves (SER) for the PFA years 2012 to 2016 for the southern NEAC complex (upper table) and the northern NEAC complex (lower table).

| Southern NEAC |  |  |
| :--- | :---: | :---: |
|  | 1SW Maturing | 1SW Non-maturing |
| Spawner Escapement Reserve (SER) | 715358 | 463566 |
| PFA Year | Probability of PFA meeting or exceeding SER |  |
| 2012 | 0.767 | 0.853 |
| 2013 | 0.673 | 0.756 |
| 2014 | 0.743 | 0.795 |
| 2015 | 0.753 | 0.797 |
| 2016 | 0.701 | 0.749 |
|  |  |  |
| Northern NEAC | 1SW Maturing |  |
|  | 201 014 | 1SW Non-maturing |
| Spawner Escapement Reserve (SER) | Probability of PFA meeting or exceeding SER |  |
| PFA Year | 0.995 | 1.000 |
| 2012 | 0.979 | 0.998 |
| 2013 | 0.962 | 0.992 |
| 2014 | 0.946 | 0.985 |
| 2015 | 0.946 | 0.983 |
| 2016 |  |  |

Table 3.6.1.1 Probability of Northern and Southern NEAC - 1SW and MSW stock complexes achieving their SERs independently and simultaneously for different catch options for the Faroes fishery in the 2013/14 to 2015/16 fishing seasons.

| Catch options for 2013/14 | TAC option <br> (t) | $\begin{gathered} \text { NEAC-N- } \\ \text { 1SW } \\ \hline \end{gathered}$ | NEAC-NMSW | $\begin{gathered} \text { NEAC-S- } \\ \text { 1SW } \\ \hline \end{gathered}$ | NEAC-SMSW | All complexes simultaneous |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| season: | 0 | 96.2\% | 99.8\% | 74.3\% | 75.6\% | 56.8\% |
|  | 20 | 96.2\% | 99.2\% | 74.2\% | 69.8\% | 52.7\% |
|  | 40 | 96.2\% | 98.2\% | 74.2\% | 63.9\% | 48.2\% |
|  | 60 | 96.1\% | 96.3\% | 74.1\% | 57.9\% | 43.3\% |
|  | 80 | 96.1\% | 93.4\% | 74.1\% | 52.1\% | 38.1\% |
|  | 100 | 96.1\% | 89.3\% | 74.0\% | 46.6\% | 32.9\% |
|  | 120 | 96.0\% | 84.3\% | 74.0\% | 41.7\% | 28.1\% |
|  | 140 | 96.0\% | 78.4\% | 73.9\% | 36.8\% | 23.4\% |
|  | 160 | 95.9\% | 71.6\% | 73.9\% | 32.5\% | 19.2\% |
|  | 180 | 95.9\% | 64.6\% | 73.8\% | 28.5\% | 15.4\% |
|  | 200 | 95.8\% | 57.6\% | 73.8\% | 25.0\% | 12.2\% |
|  |  |  |  |  |  |  |
| Catch options for 2014/15 season: | TAC option <br> (t) | NEAC-N1SW | NEAC-NMSW | NEAC-S1SW | NEAC-SMSW | All complexes simultaneous |
|  | - 0 | 94.6\% | 99.2\% | 75.4\% | 79.6\% | 59.0\% |
|  | 20 | 94.6\% | 98.2\% | 75.3\% | 75.3\% | 55.8\% |
|  | 40 | 94.6\% | 96.6\% | 75.3\% | 70.8\% | 52.0\% |
|  | 60 | 94.5\% | 94.2\% | 75.2\% | 66.4\% | 48.0\% |
|  | 80 | 94.4\% | 90.9\% | 75.2\% | 61.8\% | 43.6\% |
|  | 100 | 94.4\% | 86.8\% | 75.1\% | 57.3\% | 38.9\% |
|  | 120 | 94.3\% | 82.1\% | 75.1\% | 53.1\% | 34.4\% |
|  | 140 | 94.3\% | 76.8\% | 75.0\% | 49.0\% | 30.1\% |
|  | 160 | 94.3\% | 71.2\% | 75.0\% | 45.0\% | 25.9\% |
|  | 180 | 94.2\% | 65.5\% | 74.9\% | 41.5\% | 22.1\% |
|  | 200 | 94.2\% | 59.6\% | 74.9\% | 38.0\% | 18.6\% |
|  |  |  |  |  |  |  |
| Catch options for 2015/16 season: | TAC option <br> (t) | NEAC-N1SW | NEAC-NMSW | NEAC-S1SW | NEAC-SMSW | All complexes simultaneous |
|  | - 0 | 94.6\% | 98.5\% | 70.1\% | 79.7\% | 55.2\% |
|  | 20 | 94.6\% | 97.2\% | 70.1\% | 76.0\% | 52.4\% |
|  | 40 | 94.5\% | 95.1\% | 70.0\% | 72.2\% | 49.2\% |
|  | 60 | 94.5\% | 92.3\% | 70.0\% | 68.4\% | 45.6\% |
|  | 80 | 94.5\% | 89.0\% | 69.9\% | 64.6\% | 41.9\% |
|  | 100 | 94.4\% | 85.0\% | 69.9\% | 60.7\% | 38.0\% |
|  | 120 | 94.4\% | 80.6\% | 69.8\% | 57.1\% | 34.2\% |
|  | 140 | 94.3\% | 75.7\% | 69.8\% | 53.5\% | 30.4\% |
|  | 160 | 94.3\% | 70.6\% | 69.7\% | 50.0\% | 26.7\% |
|  | 180 | 94.2\% | 65.4\% | 69.7\% | 46.8\% | 23.4\% |
|  | 200 | 94.2\% | 60.4\% | 69.7\% | 43.7\% | 20.4\% |

Table 3.6.1.2 Forecast exploitation rates for 1SW and MSW salmon from Northern and Southern NEAC areas in all fisheries (assuming full catch allocations are taken) for different TAC options in the Faroes fishery in the 2013/14 to 2015/16 fishing seasons.

| Catch options for 2013/14 | TAC option <br> (t) | $\begin{gathered} \text { NEAC-N- } \\ \text { 1SW } \end{gathered}$ | NEAC-NMSW | NEAC-S-1SW | NEAC-SMSW |
| :---: | :---: | :---: | :---: | :---: | :---: |
| season: | 0 | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
|  | 20 | 0.0\% | 0.3\% | 0.0\% | 0.3\% |
|  | 40 | 0.0\% | 0.6\% | 0.0\% | 0.5\% |
|  | 60 | 0.0\% | 0.9\% | 0.0\% | 0.8\% |
|  | 80 | 0.0\% | 1.2\% | 0.0\% | 1.0\% |
|  | 100 | 0.1\% | 1.5\% | 0.0\% | 1.3\% |
|  | 120 | 0.1\% | 1.8\% | 0.0\% | 1.5\% |
|  | 140 | 0.1\% | 2.1\% | 0.0\% | 1.8\% |
|  | 160 | 0.1\% | 2.5\% | 0.0\% | 2.0\% |
|  | 180 | 0.1\% | 2.8\% | 0.1\% | 2.3\% |
|  | 200 | 0.1\% | 3.1\% | 0.1\% | 2.6\% |
| Catch options for 2014/15 season: | TAC option <br> (t) | $\begin{aligned} & \text { NEAC-N- } \\ & \text { 1SW } \end{aligned}$ | NEAC-NMSW | NEAC-S-1SW | $\begin{gathered} \text { NEAC-S- } \\ \text { MSW } \end{gathered}$ |
|  | 0 | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
|  | 20 | 0.0\% | 0.3\% | 0.0\% | 0.2\% |
|  | 40 | 0.0\% | 0.6\% | 0.0\% | 0.5\% |
|  | 60 | 0.0\% | 0.9\% | 0.0\% | 0.7\% |
|  | 80 | 0.0\% | 1.2\% | 0.0\% | 0.9\% |
|  | 100 | 0.0\% | 1.5\% | 0.0\% | 1.2\% |
|  | 120 | 0.1\% | 1.7\% | 0.0\% | 1.4\% |
|  | 140 | 0.1\% | 2.0\% | 0.0\% | 1.6\% |
|  | 160 | 0.1\% | 2.3\% | 0.0\% | 1.9\% |
|  | 180 | 0.1\% | 2.6\% | 0.0\% | 2.1\% |
|  | 200 | 0.1\% | 2.9\% | 0.1\% | 2.3\% |
| Catch options for 2015/16 season: | TAC option <br> (t) | $\begin{array}{r} \text { NEAC-N- } \\ \text { 1SW } \end{array}$ | $\begin{array}{r} \text { NEAC-N- } \\ \text { MSW } \end{array}$ | NEAC-S-1SW | $\begin{array}{r} \text { NEAC-S- } \\ \text { MSW } \end{array}$ |
|  | 0 | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
|  | 20 | 0.0\% | 0.3\% | 0.0\% | 0.2\% |
|  | 40 | 0.0\% | 0.5\% | 0.0\% | 0.5\% |
|  | 60 | 0.0\% | 0.8\% | 0.0\% | 0.7\% |
|  | 80 | 0.0\% | 1.0\% | 0.0\% | 0.9\% |
|  | 100 | 0.0\% | 1.3\% | 0.0\% | 1.2\% |
|  | 120 | 0.1\% | 1.5\% | 0.0\% | 1.4\% |
|  | 140 | 0.1\% | 1.8\% | 0.0\% | 1.7\% |
|  | 160 | 0.1\% | 2.0\% | 0.0\% | 1.9\% |
|  | 180 | 0.1\% | 2.3\% | 0.0\% | 2.1\% |
|  | 200 | 0.1\% | 2.5\% | 0.1\% | 2.4\% |

Table Probability (\%) of National NEAC - 1SW stock complexes achieving their SERs individually and simultaneously for 3.6.1.3 different catch options for the Faroes fishery in the 2013/14 to 2015/16 fishing seasons.

| Catch options for | TAC option (t) | Russia | Finland | Norway | Sweden | Iceland | Scotland | $\begin{gathered} \mathrm{N} . \\ \text { Ireland } \end{gathered}$ | Ireland | England \& Wales | France | All MUs simultaneous |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013/14 | 0 | 86.6\% | 84.7\% | 92.8\% | 97.7\% | 75.5\% | 54.5\% | 50.4\% | 57.0\% | 58.5\% | 27.8\% | 1.4\% |
| season: | 20 | 86.6\% | 84.6\% | 92.7\% | 97.7\% | 75.4\% | 54.5\% | 50.4\% | 57.0\% | 58.5\% | 27.7\% | 1.4\% |
|  | 40 | 86.5\% | 84.5\% | 92.6\% | 97.6\% | 75.3\% | 54.5\% | 50.3\% | 57.0\% | 58.4\% | 27.7\% | 1.4\% |
|  | 60 | 86.4\% | 84.4\% | 92.6\% | 97.6\% | 75.2\% | 54.4\% | 50.2\% | 56.9\% | 58.4\% | 27.7\% | 1.4\% |
|  | 80 | 86.4\% | 84.4\% | 92.5\% | 97.5\% | 75.1\% | 54.4\% | 50.1\% | 56.9\% | 58.3\% | 27.7\% | 1.4\% |
|  | 100 | 86.3\% | 84.3\% | 92.4\% | 97.5\% | 75.1\% | 54.3\% | 50.1\% | 56.9\% | 58.3\% | 27.6\% | 1.4\% |
|  | 120 | 86.3\% | 84.3\% | 92.4\% | 97.4\% | 75.0\% | 54.3\% | 50.0\% | 56.8\% | 58.3\% | 27.6\% | 1.4\% |
|  | 140 | 86.3\% | 84.2\% | 92.3\% | 97.3\% | 74.9\% | 54.3\% | 49.9\% | 56.8\% | 58.3\% | 27.6\% | 1.4\% |
|  | 160 | 86.2\% | 84.1\% | 92.2\% | 97.3\% | 74.9\% | 54.2\% | 49.8\% | 56.8\% | 58.2\% | 27.6\% | 1.4\% |
|  | 180 | 86.2\% | 84.0\% | 92.2\% | 97.2\% | 74.8\% | 54.2\% | 49.8\% | 56.7\% | 58.2\% | 27.5\% | 1.4\% |
|  | 200 | 86.1\% | 83.9\% | 92.1\% | 97.1\% | 74.7\% | 54.2\% | 49.7\% | 56.7\% | 58.2\% | 27.5\% | 1.4\% |
| Catch options for | TAC <br> option (t) | Russia | Finland | Norway | Sweden | Iceland | Scotland | $\begin{gathered} \mathrm{N} . \\ \text { Ireland } \end{gathered}$ | Ireland | England \& Wales | France | $\begin{array}{\|c\|} \hline \text { All MUs } \\ \text { simultaneous } \\ \hline \end{array}$ |
| 2014/15 | 0 | 83.2\% | 73.8\% | 92.0\% | 96.6\% | 75.5\% | 57.7\% | 55.2\% | 54.0\% | 58.4\% | 25.8\% | 1.2\% |
| season: | 20 | 83.2\% | 73.7\% | 92.0\% | 96.6\% | 75.4\% | 57.6\% | 55.2\% | 54.0\% | 58.4\% | 25.8\% | 1.2\% |
|  | 40 | 83.1\% | 73.6\% | 91.9\% | 96.5\% | 75.3\% | 57.6\% | 55.1\% | 54.0\% | 58.4\% | 25.7\% | 1.2\% |
|  | 60 | 83.1\% | 73.4\% | 91.8\% | 96.4\% | 75.2\% | 57.5\% | 55.0\% | 54.0\% | 58.4\% | 25.7\% | 1.2\% |
|  | 80 | 83.0\% | 73.4\% | 91.7\% | 96.4\% | 75.1\% | 57.5\% | 55.0\% | 54.0\% | 58.3\% | 25.7\% | 1.2\% |
|  | 100 | 83.0\% | 73.3\% | 91.7\% | 96.3\% | 75.1\% | 57.5\% | 54.9\% | 53.9\% | 58.3\% | 25.7\% | 1.2\% |
|  | 120 | 83.0\% | 73.1\% | 91.6\% | 96.2\% | 75.0\% | 57.5\% | 54.9\% | 53.9\% | 58.3\% | 25.7\% | 1.2\% |
|  | 140 | 82.9\% | 73.1\% | 91.6\% | 96.1\% | 74.9\% | 57.5\% | 54.8\% | 53.9\% | 58.3\% | 25.6\% | 1.2\% |
|  | 160 | 82.9\% | 73.0\% | 91.5\% | 96.0\% | 74.9\% | 57.5\% | 54.8\% | 53.9\% | 58.2\% | 25.6\% | 1.2\% |
|  | 180 | 82.8\% | 72.9\% | 91.5\% | 96.0\% | 74.8\% | 57.4\% | 54.7\% | 53.9\% | 58.2\% | 25.6\% | 1.2\% |
|  | 200 | 82.8\% | 72.8\% | 91.4\% | 95.9\% | 74.7\% | 57.4\% | 54.7\% | 53.9\% | 58.2\% | 25.5\% | 1.2\% |


| Catch options for | TAC <br> option (t) | Russia | Finland | Norway | Sweden | Iceland | Scotland | $\begin{gathered} \mathrm{N} . \\ \text { Ireland } \end{gathered}$ | Ireland | England \& Wales | France | All MUs <br> simultaneous |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015/16 | 0 | 83.0\% | 72.7\% | 92.0\% | 96.2\% | 75.5\% | 54.8\% | 61.6\% | 52.9\% | 51.2\% | 26.3\% | 1.1\% |
|  | 20 | 82.9\% | 72.6\% | 92.0\% | 96.1\% | 75.4\% | 54.8\% | 61.6\% | 52.9\% | 51.2\% | 26.2\% | 1.1\% |
|  | 40 | 82.9\% | 72.5\% | 92.0\% | 96.0\% | 75.3\% | 54.8\% | 61.5\% | 52.9\% | 51.2\% | 26.2\% | 1.1\% |
|  | 60 | 82.8\% | 72.5\% | 91.9\% | 96.0\% | 75.2\% | 54.7\% | 61.5\% | 52.9\% | 51.1\% | 26.2\% | 1.1\% |
|  | 80 | 82.8\% | 72.3\% | 91.9\% | 95.9\% | 75.1\% | 54.7\% | 61.4\% | 52.8\% | 51.1\% | 26.2\% | 1.1\% |
|  | 100 | 82.7\% | 72.2\% | 91.8\% | 95.8\% | 75.1\% | 54.7\% | 61.4\% | 52.8\% | 51.1\% | 26.1\% | 1.1\% |
|  | 120 | 82.7\% | 72.1\% | 91.8\% | 95.7\% | 75.0\% | 54.7\% | 61.3\% | 52.8\% | 51.1\% | 26.1\% | 1.1\% |
|  | 140 | 82.6\% | 72.1\% | 91.8\% | 95.7\% | 74.9\% | 54.7\% | 61.3\% | 52.8\% | 51.1\% | 26.1\% | 1.1\% |
|  | 160 | 82.6\% | 72.0\% | 91.7\% | 95.6\% | 74.9\% | 54.6\% | 61.2\% | 52.7\% | 51.0\% | 26.1\% | 1.1\% |
|  | 180 | 82.5\% | 71.9\% | 91.7\% | 95.5\% | 74.8\% | 54.6\% | 61.2\% | 52.7\% | 51.0\% | 26.1\% | 1.1\% |
|  | 200 | 82.5\% | 71.8\% | 91.6\% | 95.4\% | 74.7\% | 54.6\% | 61.1\% | 52.7\% | 51.0\% | 26.1\% | 1.1\% |

Table 3.6.1.4 Probability (\%) of National NEAC - MSW stock complexes achieving their SERs individually and simultaneously for different catch options for the Faroes fishery in the 2013/14 to 2015/16 fishing seasons.

| Catch <br> options for <br> 2013/14 <br> season: | TAC <br> option (t) | Russia | Finland | Norway | Sweden | Iceland | Scotland | N. <br> Ireland | Ireland | England <br> \& Wales | France | All MUs <br> simultaneous |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{2 0}$ | $78 \%$ | $69 \%$ | $77 \%$ | $99 \%$ | $100 \%$ | $100 \%$ | $72 \%$ | $88 \%$ | $28 \%$ | $85 \%$ | $57 \%$ |
|  | $\mathbf{4 0}$ | $60 \%$ | $73 \%$ | $97 \%$ | $99 \%$ | $100 \%$ | $63 \%$ | $77 \%$ | $26 \%$ | $82 \%$ | $54 \%$ | $2.3 \%$ |
|  | $\mathbf{6 0}$ | $52 \%$ | $69 \%$ | $96 \%$ | $98 \%$ | $99 \%$ | $59 \%$ | $73 \%$ | $25 \%$ | $81 \%$ | $52 \%$ | $1.5 \%$ |
|  | $\mathbf{8 0}$ | $44 \%$ | $66 \%$ | $94 \%$ | $97 \%$ | $99 \%$ | $55 \%$ | $68 \%$ | $24 \%$ | $80 \%$ | $51 \%$ | $0.9 \%$ |
|  | $\mathbf{1 0 0}$ | $37 \%$ | $62 \%$ | $92 \%$ | $96 \%$ | $98 \%$ | $51 \%$ | $64 \%$ | $23 \%$ | $78 \%$ | $50 \%$ | $0.5 \%$ |
|  | $\mathbf{1 2 0}$ | $31 \%$ | $59 \%$ | $89 \%$ | $95 \%$ | $98 \%$ | $48 \%$ | $61 \%$ | $22 \%$ | $77 \%$ | $48 \%$ | $0.4 \%$ |
|  | $\mathbf{1 4 0}$ | $25 \%$ | $56 \%$ | $86 \%$ | $93 \%$ | $97 \%$ | $44 \%$ | $58 \%$ | $21 \%$ | $75 \%$ | $47 \%$ | $0.2 \%$ |
|  | $\mathbf{1 6 0}$ | $21 \%$ | $54 \%$ | $83 \%$ | $92 \%$ | $95 \%$ | $40 \%$ | $55 \%$ | $20 \%$ | $74 \%$ | $45 \%$ | $0.2 \%$ |
|  | $\mathbf{1 8 0}$ | $17 \%$ | $51 \%$ | $80 \%$ | $90 \%$ | $94 \%$ | $37 \%$ | $52 \%$ | $20 \%$ | $73 \%$ | $44 \%$ | $0.1 \%$ |
|  | $\mathbf{2 0 0}$ | $14 \%$ | $48 \%$ | $77 \%$ | $88 \%$ | $93 \%$ | $34 \%$ | $50 \%$ | $19 \%$ | $71 \%$ | $43 \%$ | $0.1 \%$ |


| Catch <br> options for | TAC <br> option (t) | Russia | Finland | Norway | Sweden | Iceland | Scotland | N. <br> Ireland | Ireland | England <br> \& Wales | France | All MUs <br> simultaneous |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{2 0 1 4 / \mathbf { 5 }}$ | $\mathbf{0}$ | $75 \%$ | $69 \%$ | $99 \%$ | $100 \%$ | $100 \%$ | $73 \%$ | $87 \%$ | $29 \%$ | $82 \%$ | $52 \%$ |
| season: | $\mathbf{2 0}$ | $67 \%$ | $64 \%$ | $98 \%$ | $99 \%$ | $100 \%$ | $70 \%$ | $82 \%$ | $28 \%$ | $81 \%$ | $50 \%$ | $2.9 \%$ |
|  | $\mathbf{4 0}$ | $59 \%$ | $60 \%$ | $97 \%$ | $98 \%$ | $100 \%$ | $66 \%$ | $78 \%$ | $27 \%$ | $80 \%$ | $49 \%$ | $1.8 \%$ |
|  | $\mathbf{6 0}$ | $51 \%$ | $56 \%$ | $95 \%$ | $97 \%$ | $99 \%$ | $63 \%$ | $74 \%$ | $26 \%$ | $78 \%$ | $47 \%$ | $1.2 \%$ |
|  | $\mathbf{8 0}$ | $44 \%$ | $53 \%$ | $93 \%$ | $96 \%$ | $99 \%$ | $59 \%$ | $71 \%$ | $25 \%$ | $77 \%$ | $46 \%$ | $0.8 \%$ |
|  | $\mathbf{1 0 0}$ | $38 \%$ | $50 \%$ | $91 \%$ | $95 \%$ | $98 \%$ | $56 \%$ | $68 \%$ | $25 \%$ | $76 \%$ | $45 \%$ | $0.5 \%$ |
|  | $\mathbf{1 2 0}$ | $32 \%$ | $47 \%$ | $89 \%$ | $94 \%$ | $98 \%$ | $53 \%$ | $65 \%$ | $24 \%$ | $75 \%$ | $44 \%$ | $0.3 \%$ |
|  | $\mathbf{1 4 0}$ | $27 \%$ | $44 \%$ | $87 \%$ | $92 \%$ | $97 \%$ | $50 \%$ | $62 \%$ | $23 \%$ | $74 \%$ | $43 \%$ | $0.2 \%$ |
|  | $\mathbf{1 6 0}$ | $23 \%$ | $41 \%$ | $84 \%$ | $90 \%$ | $95 \%$ | $47 \%$ | $60 \%$ | $22 \%$ | $72 \%$ | $42 \%$ | $0.1 \%$ |
|  | $\mathbf{1 8 0}$ | $20 \%$ | $39 \%$ | $81 \%$ | $89 \%$ | $94 \%$ | $44 \%$ | $57 \%$ | $22 \%$ | $71 \%$ | $41 \%$ | $0.1 \%$ |
|  | $\mathbf{2 0 0}$ | $17 \%$ | $37 \%$ | $78 \%$ | $87 \%$ | $93 \%$ | $41 \%$ | $55 \%$ | $21 \%$ | $70 \%$ | $40 \%$ | $0.0 \%$ |


| Catch options for 2015/16 season: | TAC <br> option (t) | Russia | Finland | Norway | Sweden | Iceland | Scotland | $\begin{gathered} \mathrm{N} . \\ \text { Ireland } \end{gathered}$ | Ireland | England \& Wales | France | All MUs simultaneous |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 75\% | 68\% | 99\% | 100\% | 100\% | 69\% | 88\% | 30\% | 75\% | 50\% | 3.3\% |
|  | 20 | 68\% | 64\% | 98\% | 99\% | 100\% | 65\% | 84\% | 29\% | 74\% | 49\% | 2.2\% |
|  | 40 | 61\% | 60\% | 97\% | 98\% | 100\% | 62\% | 80\% | 28\% | 72\% | 47\% | 1.5\% |
|  | 60 | 55\% | 57\% | 95\% | 97\% | 99\% | 59\% | 77\% | 27\% | 71\% | 46\% | 1.0\% |
|  | 80 | 48\% | 54\% | 94\% | 96\% | 99\% | 56\% | 74\% | 26\% | 70\% | 45\% | 0.7\% |
|  | 100 | 43\% | 51\% | 92\% | 95\% | 98\% | 53\% | 71\% | 26\% | 68\% | 44\% | 0.5\% |
|  | 120 | 38\% | 48\% | 90\% | 93\% | 98\% | 50\% | 69\% | 25\% | 67\% | 43\% | 0.3\% |
|  | 140 | 33\% | 46\% | 88\% | 92\% | 97\% | 47\% | 67\% | 24\% | 66\% | 42\% | 0.2\% |
|  | 160 | 29\% | 44\% | 86\% | 90\% | 95\% | 44\% | 64\% | 24\% | 65\% | 41\% | 0.2\% |
|  | 180 | 25\% | 41\% | 84\% | 89\% | 94\% | 41\% | 63\% | 23\% | 63\% | 40\% | 0.1\% |
|  | 200 | 22\% | 40\% | 82\% | 87\% | 93\% | 39\% | 61\% | 22\% | 62\% | 39\% | 0.1\% |

Table 3.6.1.5. Compliance with river-specific conservation limits for individual river stocks, before homewater fisheries, within each jurisdiction in the NEAC area in 2012 (except Norway where data are for 2011). NA = not available.

| Country | No. rivers | $\begin{aligned} & \text { No. with } \\ & \text { CL } \end{aligned}$ | No. ASESSED <br> FOR <br> COMPLIANCE | No. <br> ATTAINING CL | \% ATtAINING CL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Northern NEAC |  |  |  |  |  |
| Russia | 112 | 80 | 7 | 6 | 86 |
| Finland/Norway (Tana/Teno) | 1 | 1 | NA | NA | NA |
| Norway (2011) | $481$ | 439 | 186 | 163 | 88 |
| Sweden | 23 | 17 | NA | NA | NA |
| Iceland | 100 | 0 | NA | NA | NA |
| Southern NEAC |  |  |  |  |  |
| UK Scotland | 383 | 0 | 0 | NA | NA |
| UK N. Ireland | 15 | 10 | NA | NA | NA |
| UK England/Wales | 78 | 64 | NA | NA | NA |
| Ireland | 141 | 141 | 141 | 58 | 41 |
| France (MSW) | 35 | 31 | NA | NA | NA |



Figure 3.1.3.1. Overview of effort as reported for various fisheries and countries 1971-2012 in the northern NEAC area.


Figure 3.1.3.2. Overview of effort as reported for various fisheries and countries 1971-2012 in the southern NEAC area.


Figure 3.1.4.1. Nominal catches of salmon and 5-year running means in the southern and northern NEAC areas, 1971-2012.


Figure 3.1.5.1. Proportional change (\%) over years in cpue estimates in various rod and net fisheries in northern and southern NEAC area (* Information on cpue in Barents Sea rivers in Russia to 2008).


Figure 3.1.6.1. Percentage of 1SW salmon in the reported catch for northern NEAC countries, 19872012.


Figure 3.1.6.2. Percentage of 1SW salmon in the reported catch for southern NEAC countries, 19872012.


Figure 3.1.9.1. Mean annual exploitation rate of wild 1SW and MSW salmon by combined commercial and recreational fisheries in northern NEAC countries from 1983 to 2012.


Figure 3.1.9.2. Mean annual exploitation rate of wild 1SW and MSW salmon by combined commercial and recreational fisheries in southern NEAC countries from 1971 to 2012.


Figure 3.1.9.3. The rate of change of exploitation of 1SW and MSW salmon in northern NEAC (left) and southern NEAC (right) countries.


Figure 3.3.4.1a. Summary of fisheries and stock description, River Teno/Tana (Finland and Norway combined).


Figure 3.3.4.1b. Summary of fisheries and stock description, France. The national CL analysis is shown for information only. A river-specific CL is used for the assessment.


Figure 3.3.4.1c. Summary of fisheries and stock description, Iceland.


Figure 3.3.4.1d. Summary of fisheries and stock description, Ireland. The national CL analysis is shown for information only. A river-specific CL is used for the assessment.


Figure 3.3.4.1e. Summary of fisheries and stock description, Norway (minus Norwegian catches from the R. Teno/Tana). The regional CL analyses is shown for information only. A river-specific CL is used for the assessment.


Figure 3.3.4.1f. Summary of fisheries and stock description, Russia.


Figure 3.3.4.1g. Summary of fisheries and stock description, Sweden.


Figure 3.3.4.1h. Summary of fisheries and stock description, UK(England \& Wales). The national CL analysis is shown for information only. A river-specific CL is used for the assessment.


Figure 3.3.4.1i. Summary of fisheries and stock description, UK(N.Ireland).


Figure 3.3.4.1j. Summary of fisheries and stock description, UK(Scotland).


Figure 3.3.5.1. Estimated PFA (left panels) and spawning escapement (right panels) with $95 \%$ confidence limits, for maturing 1SW (1SW spawners) and non-maturing 1SW (MSW spawners) salmon in northern (NEAC-N) and southern (NEAC-S) NEAC stock complexes.


Figure 3.3.7.1. Comparison of the percent change in the five year mean return rates for 1SW and 2SW wild salmon smolts to rivers of northern (top) and southern NEAC (bottom) areas for the 2002 to 2006 and 2007 to 2011 smolt years ( 2001 to 2005 and 2006 to 2010 for 2SW salmon). Filled circles are for 1SW and open circles are for 2SW dataseries. Triangles indicate all ages without separation into 1SW and 2SW smolts. Populations with at least three datapoints in each of the two time periods are included in the analysis. The scale of change in some rivers is influenced by low return numbers, where a few fish more or less returning may have a significant impact on the percent change.


Figure 3.3.7.2. Comparison of the percent change in the five-year mean return rates for 1SW and 2SW hatchery salmon smolts to rivers of northern (top) and Southern NEAC (bottom) areas for the 2002 to 2006 and 2007 to 2011 smolt years ( 2001 to 2005 and 2006 to 2010 for 2SW salmon). Filled circles are for 1SW and open circles are for 2SW dataseries. Triangles indicate all ages without separation into 1SW and 2SW smolts. Populations with at least three datapoints in each of the two time periods are included in the analysis. The scale of change in some rivers is influenced by low return numbers, where a few fish more or less returning may have a significant impact on the percent change.


## Year of smolt migration

Figure 3.3.7.3. Standardized mean (one standard error bars) annual return rates of wild (left hand panels) and hatchery origin (right hand panels) smolts to 1SW and 2SW salmon to northern and southern areas. The standardized values are annual means derived from a general linear model analysis of rivers in a region. Survival rates were $\log$ transformed prior to analysis. Note $y$-scale differences among panels.


Figure 3.4.3.1. Probability of simultaneous attainment of CLs for different numbers of management units with a $95 \%$ chance of attainment in each management unit independently.


Figure 3.5.2.1. Southern NEAC PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1978 to 2016. The last five years ( 2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the age-specific SER values. Box and whiskers show the 5th, 25th, 50th, 75th and 95th Bayesian credibility intervals (BCIs).


Figure 3.5.2.2. Northern NEAC PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1991 to 2016. The last five years ( 2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the age-specific SER values. Box and whiskers show the 5th, 25th, 50th, 75th and 95th Bayesian credibility intervals (BCIs).


Probability that PFAs will be Greater than or
equal to country and age specific SERs

|  |  | Maturing |
| :--- | :--- | :--- |
| France | SER | 22120 |
| Year | $p$ | 8493 |
| 2012 | 0.079 | $p$ |
| 2013 | 0.181 | 0.324 |
| 2014 | 0.274 | 0.568 |
| 2015 | 0.255 | 0.512 |
| 2016 | 0.255 | 0.490 |

Figure 3.5.4.1. France: PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1978 to 2016. The last five years (2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the agespecific SER values. Box and whiskers show the 5th, 25th, 50th, 75th and 95th BCIs.







Probability that PFAs will be Greater than or
equal to country and age specific SERs

|  |  | Maturing |
| :--- | :--- | :--- |
| Ireland | SER | 268832 |
| Year | $p$ | 78174 |
| 2012 | 0.631 | $p$ |
| 2013 | 0.567 | 0.188 |
| 2014 | 0.566 | 0.232 |
| 2015 | 0.538 | 0.281 |
| 2016 | 0.521 | 0.290 |

Figure 3.5.4.2. Ireland: PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1978 to 2016. The last five years (2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the agespecific SER values. Box and whiskers show the 5th, 25th, 50th, 75th and 95th BCIs.







Probability that PFAs will be Greater than or
equal to country and age specific SERs
$\left.\begin{array}{lll}\hline & & \text { Maturing } \\ \hline \text { UK (N. Ireland) } & \text { SER } & 23850 \\ \hline \text { Year } & & p\end{array}\right) 3093$

Figure 3.5.4.3. UK(N. Ireland): PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1978 to 2016. The last five years (2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the age-specific SER values. Box and whiskers show the 5th, 25th, 50th, 75th and 95th BCIs.







Probability that PFAs will be Greater than or
equal to country and age specific SERs

|  |  | Maturing |
| :--- | :--- | :--- |
| UK (England \& Wales) | SER | 69272 |
| Year | $p$ | 50802 |
| 2012 | 0.505 | $p$ |
| 2013 | 0.438 | 0.871 |
| 2014 | 0.583 | 0.783 |
| 2015 | 0.582 | 0.840 |
| 2016 | 0.508 | 0.819 |

Figure 3.5.4.4. UK(England \& Wales): PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1978 to 2016. The last five years ( 2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the age-specific SER values. Box and whiskers show the 5th, 25th, 50th, 75 th and 95 th BCIs.



| Probability that PFAs will be Greater than or |  |  |
| :--- | :--- | :--- |
| equal to country and age specific SERs |  |  |
|  |  | Maturing |
| UK (Scotland) | SER | 305206 |
| Year | $p$ | 320577 |
| 2012 | 0.507 | $P$ |
| 2013 | 0.485 | 0.790 |
| 2014 | 0.541 | 0.706 |
| 2015 | 0.573 | 0.718 |
| 2016 | 0.543 | 0.729 |

Figure 3.5.4.5. UK(Scotland): PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1978 to 2016. The last five years ( 2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the age-specific SER values. Box and whiskers show the 5th, 25th, 50th, 75th and 95th BCIs.







Probability that PFAs will be Greater than or
equal to country and age specific SERs

|  | Maturing | Non-maturing |
| :--- | :--- | :--- |
| Iceland - SW | SER | 25692 |
| Year | $p$ | 2192 |
| 2012 | 0.988 | $P$ |
| 2013 | 0.974 | 0.949 |
| 2014 | 0.965 | 0.915 |
| 2015 | 0.952 | 0.894 |
| 2016 | 0.902 | 0.815 |

Figure 3.5.4.6. Iceland (south/west regions): PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1978 to 2016. The last five years ( 2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the age-specific SER values. Box and whiskers show the 5th, 25th, 50th, 75th and 95th BCIs.







Probability that PFAs will be Greater than or
equal to country and age specific SERs

|  | Maturing | Non-maturing |
| :--- | :--- | :--- |
| Russia | SER | 84628 |
| Year | $p$ | 73894 |
| 2012 | 0.956 | $P$ |
| 2013 | 0.880 | 0.887 |
| 2014 | 0.868 | 0.780 |
| 2015 | 0.835 | 0.772 |
| 2016 | 0.832 | 0.745 |

Figure 3.5.4.7. Russia: PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1991 to 2016. The last five years (2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the agespecific SER values. Box and whiskers show the 5th, 25th, 50th, 75th and 95th BCIs.







Probability that PFAs will be Greater than or
equal to country and age specific SERs

|  |  | Maturing |
| :--- | :--- | :--- |
| Finland | SER | 16662 |
| Year | $p$ | 27772 |
| 2012 | 0.947 | $P$ |
| 2013 | 0.919 | 0.920 |
| 2014 | 0.846 | 0.890 |
| 2015 | 0.738 | 0.809 |
| 2016 | 0.728 | 0.686 |

Figure 3.5.4.8. Finland: PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1991 to 2016. The last five years (2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the age-specific SER values. Box and whiskers show the 5th, 25th, 50th, 75th and 95th BCIs.







Probability that PFAs will be Greater than or
equal to country and age specific SERs

|  | Maturing | Non-maturing |
| :--- | :--- | :--- |
| Norway | SER | 89774 |
| 116367 |  |  |
| Year | $p$ | $P$ |
| 2012 | 0.972 | 0.999 |
| 2013 | 0.951 | 0.997 |
| 2014 | 0.926 | 0.992 |
| 2015 | 0.918 | 0.987 |
| 2016 | 0.920 | 0.986 |

Figure 3.5.4.9. Norway: PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1991 to 2016. The last five years (2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the age-specific SER values. Box and whiskers show the 5th, 25th, 50 th, 75 th and 95 th BCIs.







Probability that PFAs will be Greater than or
equal to country and age specific SERs

|  |  | Maturing |
| :--- | :--- | :--- |
| Sweden | SER | 1741 |
| Year | p | 2169 |
| 2012 | 0.996 | p |
| 2013 | 0.988 | 1.000 |
| 2014 | 0.977 | 1.000 |
| 2015 | 0.967 | 0.999 |
| 2016 | 0.963 | 0.997 |

Figure 3.5.4.10. Sweden: PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1991 to 2016. The last five years (2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the age-specific SER values. Box and whiskers show the 5th, 25th, 50th, 75th and 95th BCIs.







Probability that PFAs will be greater than or
equal to country and age specific SERs
$\left.\begin{array}{lll}\hline & & \text { Maturing } \\ \hline \text { Iceland-NE } & \text { SER } & \mathbf{8 2 0 9}\end{array}\right)$ Non-maturing

Figure 3.5.4.11. Iceland (north/east regions): PFA maturing and non-maturing, lagged eggs from 1SW and MSW, proportion 1SW maturing, and the productivity parameter values for PFA years 1991 to 2016. The last five years ( 2012 to 2016) are forecasts in all cases. The dashed horizontal lines in the upper panels are the age-specific SER values. Box and whiskers show the 5th, 25th, 50th, 75 th and 95th BCIs.


Figure 3.6.1.1. Probability of northern and southern NEAC - 1SW and MSW stock complexes, and all stock complexes simultaneously, achieving their SERs for different catch options for the Faroes fishery in the 2013/14 to 2015/16 fishing seasons.


Figure 3.7.2.1. Number of retained indicators within each stock complex as a function of the $\mathbf{R}^{2}$ criterion chosen to select the indicators for the Faroes FWI.

## 4 North American commission

The previous advice provided by ICES (2009a) indicated that there were no mixed-stock fishery catch options on the 1SW non-maturing salmon component for the 2009 to 2011 PFA years, and this year's assessment confirms that advice. The NASCO Framework of Indicators of North American stocks for 2012 did not indicate the need for a revised analysis of catch options and no new management advice for 2013 is provided. The assessment was updated to 2012 and the stock status is consistent with the previous years' assessments and catch advice.

### 4.1 NASCO has requested ICES to describe the key events of the 2012 fisheries

### 4.1.1 Key events of the 2012 fisheries

- There were no new significant events reported for 2012 in the NAC area.
- The majority of harvest fisheries were directed to small salmon.
- The 2012 provisional harvest was 134.7 t, comprised of 46891 small salmon and 11671 large salmon, $27 \%$ fewer small salmon and $15 \%$ fewer large salmon compared to 2011.
- Overall, catches remain very low relative to pre-1990 values.


### 4.1.2 Gear and effort

## Canada

The 23 areas for which the Department of Fisheries and Oceans (DFO) manages the salmon fisheries are called Salmon Fishing Areas (SFAs); for Quebec, the management is delegated to the province of Quebec (Ministère du Développement durable, de l'Environnement, de la Faune et des Parcs) and the fishing areas are designated by Q1 through Q11 (Figure 4.1.2.1). Harvests (fish which were retained) and catches (including harvests and fish caught-and-released in recreational fisheries) are categorized in two size groups: small and large. Small salmon, generally 1SW, in the recreational fisheries refer to salmon less than 63 cm fork length, whereas in commercial fisheries, it refers to salmon less than 2.7 kg whole weight. Large salmon, generally MSW, in recreational fisheries are greater than or equal to 63 cm fork length and in commercial fisheries refer to salmon greater than or equal to 2.7 kg whole weight.

Three groups exploited salmon in Canada in 2012; Aboriginal peoples, residents fishing for food in Labrador, and recreational fishers. There were no commercial fisheries in Canada in 2012.

In 2012, four subsistence fisheries harvested salmon in Labrador: 1) Nunatsiavut Government (NG) members fishing in the northern Labrador communities of Rigolet, Makkovik, Hopedale, Postville, and Nain and in Lake Melville; 2) Innu Nation members fishing in Natuashish and in Lake Melville from the community of Sheshatshiu; 3) NunatuKavut Community Council (NCC) members fishing in southern Labrador from Fish Cove Point to Cape St Charles and, 4) Labrador residents fishing in Lake Melville and various coastal communities. The NG, Innu, and NCC fisheries were monitored by

Aboriginal Fishery Guardians jointly administered by the Aboriginal groups and the DFO, as well as, by DFO Fishery Officers and Guardian staff. The Nunatsiavut Government is directly responsible through the Torngat Fisheries Board for regulating its fishery through its Conservation Officers. The fishing gear is multifilament gillnets of 15 fathoms $(27.4 \mathrm{~m})$ in length of a stretched mesh size ranging from 3 to 4 inches ( 7.6 to 10.2 cm ). Although nets are mainly set in estuarine waters some nets are also set in coastal areas usually within bays. Catch statistics are based on logbook reports.

Most catches (95\%, Figure 2.1.1.2) in Canada now take place in rivers or in estuaries. Fisheries are principally managed on a river-by-river basis and, in areas where retention of large salmon is allowed, are closely controlled. The commercial fisheries are now closed and the remaining coastal food fisheries in Labrador are mainly located in bays generally inside the headlands. Sampling of this fishery occurred in 2012 with the intent of using genetic markers to identify the origin of harvested salmon.

The following management measures were in effect in 2012.

## Aboriginal peoples' food, social, and ceremonial (FSC) fisheries

In Quebec, Aboriginal peoples' fisheries took place subject to agreements or through permits issued to the bands. There are ten bands with subsistence fisheries in addition to the fishing activities of the Inuit in Ungava (Q11), who fished in estuaries or within rivers. The permits generally stipulate gear, season, and catch limits. Catches have to be reported collectively by each Aboriginal user group. However, if reports are not available, the catches are estimated. In the Maritimes (SFAs 15 to 23), FSC agreements were signed with several Aboriginal peoples' groups (mostly First Nations) in 2012. The signed agreements often included allocations of small and large salmon and the area of fishing was usually in-river or estuaries. Harvests that occurred both within and outside agreements were obtained directly from the Aboriginal peoples. In Labrador (SFAs 1 and 2), fishery arrangements with the Nunatsiavut Government, the Innu First Nation, and the NCC, resulted in fisheries in estuaries and coastal areas. By agreement with First Nations, there were no FSC fisheries for salmon on the island of Newfoundland in 2012. Harvest by Aboriginal peoples with recreational licences is reported under the recreational harvest categories.

## Resident food fisheries in Labrador

The DFO is responsible for regulating the Resident Fishery. In 2012, a licensed subsistence trout fishery for local residents took place, using gillnets, in Lake Melville (SFA 1) and in estuary and coastal areas of Labrador (SFA 1 and 2). Residents who requested a licence were permitted to retain a seasonal bycatch of three salmon of any size while fishing for trout and charr; three salmon tags accompanied each licence. When the bycatch of three salmon was caught the resident fishers were required to remove their net from the water. If bycatch during a single gillnet set exceeded three salmon, resident fishers were required to discard the excess fish. All licensees were requested to complete logbooks.

## Recreational fisheries

Licences are required for all persons fishing recreationally for Atlantic salmon. Gear is restricted to fly fishing and there are daily and seasonal bag limits. Recreational fisheries management in 2012 varied by area and large portions of the southern areas remained
closed to all directed salmon fisheries (Figure 4.1.2.2). Large salmon were no longer permitted to be retained in Labrador as of 2011. Except for 42 rivers in Quebec, only small salmon could be retained in the recreational fisheries.

In 2012, on Prince Edward Island, there was no directed Atlantic salmon recreational fishery as no recreational fishery licence for Atlantic salmon was issued by the province of Prince Edward Island. The recreational catch of Atlantic salmon for Prince Edward Island for 2012 is recorded as zero. As in other areas of eastern Canada, there is no estimate of salmon released as bycatch in non-salmon directed recreational fisheries.

## USA

There were no recreational or commercial fisheries for anadromous Atlantic salmon in the USA in 2012.

## France (Islands of Saint-Pierre et Miquelon)

Nine professional and 60 recreational gillnet licences were issued in 2012, an increase of four recreational licences from 2011. Professional licences have a maximum authorization of three nets of 360 metres maximum length whereas the recreational licence is restricted to one net of 180 metres. The time-series of available effort data (licence numbers) is in Table 4.1.2.1.

### 4.1.3 Catches in 2012

## Canada

The provisional harvest of salmon in 2012 by all users was 134.7 t , about $25 \%$ lower than the 2011 harvest of 179 t (Table 2.1.1.2; Figure 4.1.3.1). The 2012 harvest was 46981 small salmon and 11671 large salmon, $27 \%$ fewer small salmon and $15 \%$ fewer large salmon compared to 2011. This is the third lowest catch in the time-series since 1960:, there has been a dramatic decline in harvested tonnage since 1988 in large part the result of the reductions in commercial fisheries effort; the closure of the insular Newfoundland commercial fishery in 1992, the closure of the Labrador commercial fishery in 1998 and the closure of the Quebec commercial fishery in 2000.

## Aboriginal peoples' FSC fisheries

The total harvest by Aboriginal people in 2012 was 60.5 t (Table 4.1.3.1). Harvest (by weight) decreased by $14 \%$ from 2011. The reported catch in 2012 was the fifth highest value in the time-series.

## Residents fishing for food in Labrador

The estimated catch for the fishery in 2012 was 1.7 t . This represents approximately 650 fish, $32 \%$ of which were large (Table 4.1.3.2).

## Recreational fisheries

Harvest in recreational fisheries in 2012 totalled 37720 small and large salmon (approximately 72 t ), decreased $30 \%$ from the 2011 harvest level and $14 \%$ from the previous fiveyear average and remains at low levels similar to the previous decade (Table 4.1.3.3; Fig-
ure 4.1.3.2). The small salmon harvest of 35040 fish was $30 \%$ lower than the 2011 harvest. The large salmon harvest of 2680 fish was $35 \%$ lower than the 2011 harvest and occurred only in Quebec. The small salmon size group has contributed $89 \%$ on average of the total recreational harvests since the imposition of catch-and-release recreational fisheries in the Maritimes and insular Newfoundland (SFA 3 to 14B, 15 to 23) in 1984. In 2012, approximately 50800 salmon (about 32500 small and 18300 large) were caught and released (Table 4.1.3.4), representing about $57 \%$ of the total number caught (including retained fish). For Prince Edward Island in 2012, there were no recreational fishery salmon licences issued and hence no catch-and-release values to report. There is some mortality on these released fish, which is accounted for in the spawner estimates.

Recreational catch statistics for Atlantic salmon are not collected regularly in Canada and there is no mechanism in place that requires anglers to report their catch statistics, except in Quebec. The last recreational angler survey for New Brunswick was conducted in 1997 and the catch rates for the Miramichi from that survey have been used to estimate catches (both harvest and catch-and-release) for all subsequent years, except for 2011 and 2012 where there is no estimate made for release of salmon kelts. The reliability of recreational catch statistics could be improved in all areas of Canada.

## Commercial fisheries

All commercial fisheries for Atlantic salmon remained closed in Canada in 2012 and the catch therefore was zero.

## Unreported catches

The unreported catch estimate for Canada is complete and totalled 30.5 t in 2012, a value similar to that reported for 2011. The majority of this unreported catch is illegal fisheries directed at salmon (Tables 2.1.3.1, 2.1.3.2). Of the unreported catch which could be attributed to a geographic location $(14.7 \mathrm{t}), 9.4 \mathrm{t}$ was considered to have occurred in inland waters and 5.4 t in tidal waters.

## USA

There are no commercial or recreational fisheries for Atlantic salmon in USA and the catch therefore was zero. Unreported catches in the USA were estimated to be 0 t .

## France (Islands of Saint-Pierre et Miquelon)

A total harvest of 1.5 t was reported in the professional and recreational fisheries in 2012, a decrease of $61 \%$ from the 2011 reported harvest of 3.8 t , and the lowest value since 1997 (Table 4.1.2.1).

There are no unreported catch estimates.

### 4.1.4 Harvest of North American salmon, expressed as 2 SW salmon equivalents

Harvest histories (1972 to 2012) of salmon, expressed as 2 SW salmon equivalents are provided in Table 4.1.4.1. The Newfoundland-Labrador commercial fishery historically was a mixed-stock fishery and harvested both maturing and non-maturing 1 SW salmon as well as 2SW maturing salmon. The harvest in these fisheries of repeat spawners and older sea ages was not considered in the run reconstructions.

Harvests of 1SW non-maturing salmon in Newfoundland-Labrador commercial fisheries have been adjusted by natural mortalities of $3 \%$ per month for 13 months, and 2SW harvests in these same fisheries have been adjusted by one month to express all harvests as 2SW equivalents in the year and time they would reach rivers of origin. The Labrador commercial fishery has been closed since 1998. Harvests from the Aboriginal Peoples' fisheries in Labrador (since 1998) and the residents' food fishery in Labrador (since 2000) are both included. Mortalities in mixed-stock and terminal fisheries areas in Canada were summed with those of USA to estimate total 2SW equivalent mortalities in North America. The terminal fisheries included coastal, estuarine and river catches of all areas, except Newfoundland and Labrador where only river catches were included and excluding Saint-Pierre et Miquelon. Harvest equivalents within North America peaked at about 363000 in 1976 and have remained below 14000 2SW salmon equivalents since 1999 (Таble 4.1.4.1).

In the most recent year, the harvest of cohorts destined to be 2 SW salmon in terminal fisheries of North America was $64 \%$ of the total North American catch of 2SW salmon. The percentages of harvests occurring in terminal fisheries ranged from 19 to $32 \%$ during 1972 to 1990 and 61 to $89 \%$ during 1993 to 2012 (Table 4.1.4.1). Percentages increased significantly since 1992 with the reduction and closures of the Newfoundland and Labrador commercial mixed-stock fisheries.

### 4.1.5 Origin and composition of catches

In the past, salmon from both Canada and the USA were taken in the commercial fisheries of eastern Canada. The Aboriginal Peoples' and resident food fisheries that occur in Labrador may intercept salmon from other areas of North America. However, in 2009 to 2012, there were no reports of tagged salmon from other areas in these fisheries. Also none of the salmon sampled during the Food Fishery Sampling Programme in those years were carrying a tag or mark. No tags were reported from the fishery in Saint-Pierre et Miquelon. No tagged salmon of USA origin were reported in Canadian fisheries in 2012.

## Results of sampling programme for Labrador subsistence fisheries

A sampling programme of the subsistence fisheries in Labrador continued in 2012, conducted by the NunatuKavut Community Council (NCC) and Conservation Officers of the Nunatsiavut Government (NG). Landed fish were sampled opportunistically for fork length, weight (gutted weight or whole weight if available) and sex. Scales were taken for age analysis and an adipose fin clip was taken for genetic analysis. Fish were also examined for the presence of external tags, brands or elastomer marks.

In 2012, a total of 420 samples were collected from the Labrador subsistence fisheries, 151 from northern Labrador (SFA 1A), 42 from Lake Melville (SFA 1B) and 227 samples from southern Labrador (SFA 2) (Figure 4.1.2.1). Based on the interpretation of the scale samples, $79 \%$ of all the samples taken were 1 SW salmon, $10 \%$ were 2 SW, and $11 \%$ were previously spawned salmon. One fish was 3SW. The majority of salmon sampled were river ages 3 to 5 years ( $96 \%$ ) (modal age 4). There were very few river age $1(0.2 \%$ ) or river age 2 (1.9\%) salmon sampled suggesting, as in previous years (2006 to 2011), that very few salmon from the most southern stocks of North America (USA, Scotia-Fundy) were exploited in these fisheries.

| Percentage of samples by river age within the three sampled areas |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | Number of Samples | River Age |  |  |  |  |  |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 |
| Northern Labrador (SFA 1A) | 149 | 0.7 | 3.4 | 19.5 | 56.4 | 18.1 | 2.0 |
| Lake Melville (SFA 1B) | 40 | 0.0 | 5.0 | 15.0 | 67.5 | 12.5 | 0.0 |
| Southern Labrador (SFA 2) | 225 | 0.0 | 0.4 | 23.6 | 54.2 | 20.4 | 1.3 |
| All areas | 414 | 0.2 | 1.9 | 21.3 | 56.3 | 18.8 | 1.4 |

A collaborative project between the DFO, the Atlantic Salmon Federation, the Nunatsiavut Government and the NunatuKavut Community Council was initiated in 2011 to examine the stock composition of the subsistence catch of salmon in Labrador. This project involved collection of genetic samples from the 2011 catch and examination of scales collected in the 2006-2010 period. Samples were also collected in 2012. Genetic analysis involved the genotyping of 15 microsatellite loci from approximately 1600 Atlantic salmon from the subsistence harvest in coastal Labrador and has recently been completed. Salmon baseline data for Newfoundland and Labrador (approximately 80 rivers) collected by DFO has been completed and will be integrated into the Canadian Atlantic salmon genetic database being produced by the collaborative research group led by Laval University. Incorporation of the USA salmon baseline into the analyses is ongoing. Standardization and integration of the regional components of the Canadian baseline are underway and once completed the analysis of fishery composition will be conducted and the project report is expected in 2013.

The Working Group noted that this sampling programme provides biological characteristics of the harvest and the origin of the fish in the fishery. These are important parameters in the Run Reconstruction Model for North America.

## Sampling programme for Saint-Pierre et Miquelon

No sampling of the salmon catches was conducted in 2012. The Working Group had previously been informed of the results of genetic stock identification of samples collected in 2010 and 2011.

The Working Group had previously identified a number of issues with the sampling programme that if accounted for, would greatly increase the value of the data (ICES, 2011b). Future genetic analyses should use the extensive baselines of Canadian and USA salmon populations.

The Working Group welcomed the past efforts to sample the catches at Saint-Pierre et Miquelon. The collaboration between French and Canadian researchers was encouraged to ensure that adequate samples are collected and that the North American genetic baseline is used in the analysis of these samples. This would address the gaps identified in the previous sampling activities (ICES, 2011b).

## Recommendations for future activities

The Working Group recommends that sampling of the Labrador and Saint-Pierre et Miquelon fisheries be continued and expanded (i.e. sample size, geographic coverage, tissue samples, seasonal distribution of the samples) in future years and analysed using the North American baseline to improve the information on biological characteristics and stock origin of salmon harvested in these mixed-stock fisheries.

### 4.1.6 Exploitation rates

## Canada

In the insular Newfoundland recreational fishery, exploitation rates for retained small salmon ranged from a high of $19 \%$ on Exploits River to a low of 5\% on Terra Nova River. Overall, exploitation rates of small salmon in these rivers declined from $30 \%$ in 1986 to approximately $14 \%$ in 2012. In Sand Hill River, Labrador, exploitation rate on small salmon was $2.6 \%$ and no large salmon were reported as retained in 2012.

In Quebec, the 2012 total fishing exploitation rate was about $18 \%$, similar to the average of the five previous years. Native peoples' fishing exploitation rate was $7 \%$ of the total return. Recreational fishing exploitation rate was $11 \%$ on the total run, $16 \%$ for the small and $7 \%$ for the large salmon, slightly lower than the previous five year average of $18 \%$ for small salmon and $8 \%$ for large salmon.

## USA

There was no exploitation of anadromous USA salmon in homewaters.

## Exploitation trends for North American salmon fisheries

Annual exploitation rates of small salmon (mostly 1SW) and large salmon (mostly MSW) in North America for the 1971 to 2012 time period were calculated by dividing annual harvests in all North American fisheries by annual estimates of the returns to North America prior to any homewater fisheries. The fisheries included coastal, estuarine and river fisheries in all areas, as well as the commercial fisheries of Newfoundland and Labrador which harvested salmon from all regions in North America.

Exploitation rates of both small and large salmon fluctuated annually but remained relatively steady until 1984 when exploitation of large salmon declined sharply with the introduction of the non-retention of large salmon in angling fisheries and reductions in commercial fisheries (Figure 4.1.5.1). Exploitation of small salmon declined steeply in North America with the closure of the Newfoundland commercial fishery in 1992. Declines continued in the 1990s with continuing management controls in all fisheries to reduce exploitation. In the last few years, exploitation rates on small salmon and large salmon have remained at the lowest in the time-series, averaging $13.5 \%$ for large salmon and $15.7 \%$ for small salmon over the past ten years. However, exploitation rates across regions within North America are highly variable.

### 4.2 Management objectives and reference points

Management objectives are described in Section 1.4.

There were no changes to the 2SW salmon Conservation Limits (CLs) from those identified previously. CLs for 2SW salmon for Canada total 123349 and for the USA, 29 199, for a combined total of 152548.

| Country and Comission Area | Stock Area | 2SW SPAWNer requirement |
| :--- | :--- | :---: |
|  | Labrador | 34746 |
|  | Newfoundland | 4022 |
|  | Gulf of St Lawrence | 30430 |
|  | Quebec | 29446 |
| Canada Total | Scotia-Fundy | 24705 |
| USA |  | 123349 |
| North American Total |  | 29199 |

### 4.3 Status of stocks

To date, 1082 Atlantic salmon rivers have been identified in eastern Canada and 21 rivers in eastern USA, where salmon are or were present within the last half century. Assessments were reported for 74 of these rivers in 2012.

### 4.3.1 Smolt abundance

## Canada

Wild smolt production was estimated in eleven rivers in 2012. Smolt production increased from 2011 in three rivers (range $19 \%$ to $95 \%$ ), decreased in three rivers (range $34 \%$ to $44 \%$ ) and remained unchanged (within $+/-10 \%$ ) in four rivers. There was no 2011 estimate for one of the eleven rivers. The relative smolt production, scaled to the size of the river using the conservation egg requirements, was highest in the Gulf region and lowest in the southern rivers of the Scotia-Fundy region (Figure 4.3.1.1). Significant linear declines in smolt production ( $\mathrm{p}<0.05$ ) have been observed in St Jean (1989-2012) and de la Trinite (1984-2012) (Quebec), whereas production has increased significantly in Western Arm Brook (Newfoundland) (1971-2012).

## USA

Wild salmon smolt production has been estimated on the Narraguagus River from 1997 to 2012 (16 years) (Figure 4.3.1.1). Smolt production in 2012 was 31\% less than in 2011 and has declined significantly since 1997 ( $\mathrm{p}<0.05$ ).

### 4.3.2 Estimates of total adult abundance by geographic area

Returns of small (1SW), large, and 2SW salmon (a subset of large) to each region (Figures 4.3.2.1, 4.3.2.2 and 4.3.2.3; and Annex 6) were originally estimated by the methods and variables developed by Rago et al. (1993) and reported by ICES (1993). The returns for individual river systems and management areas for both sea age groups were derived from a variety of methods. These methods included counts of salmon at monitoring facilities, population estimates from mark-recapture studies, and applying angling and
commercial catch statistics, angling exploitation rates, and measurements of freshwater habitat. The 2 SW component of the large returns was determined using the sea age composition of one or more indicator stocks.

Returns are the number of salmon that returned to the geographic region, including fish caught by homewater commercial fisheries, except in the case of the Newfoundland and Labrador regions where returns do not include landings in commercial and food fisheries. This avoided double counting fish because commercial catches in Newfoundland and Labrador and food fisheries in Labrador were added to the sum of regional returns to create the PFA of North American salmon.

Total returns of salmon to USA rivers are the sum of trap catches and redd based estimates.

## Canada

## Labrador

The median of the estimated returns of small salmon in 2012 to Labrador (172 800) was $36 \%$ lower than the previous year and $1 \%$ higher than the previous five year mean (170 336, Figure 4.3.2.1). The median of the estimated 2SW returns in 2012 to Labrador (22060) was $22 \%$ lower than the previous year and $17 \%$ higher than the previous five year mean (18 846, Figure 4.3.2.3).

Since 2002, Labrador regional estimates are generated from data collected at four counting facilities, one in SFA 1 and three in SFA 2 (Figure 4.1.2.1). In 2010 and 2012, only two of three facilities in SFA 2 were operational. The production area in SFA 1 is approximately equal to the production area in SFA 2. The current method to estimate Labrador returns assumes that the total returns to the northern area are represented by returns at the single monitoring facility in SFA 1 and returns in the southerly areas (SFA2 and 14b) are represented by returns at the monitoring facilities in SFA 2. Further work is needed to understand the best use of these data in describing stock status and the Working Group recommends that additional data be considered in Labrador to better estimate salmon returns in that region. Nonetheless, the changes in abundance reported for Labrador were in line with changes observed elsewhere in North America and consistent with coherent patterns operating over a broad geographic scale.

## Newfoundland

Finalized angling information from 2011 was used to update estimates of salmon returns in that year. The median of the estimated returns of small salmon in 2012 to Newfoundland (242 300) was approximately equal to the previous year and $4 \%$ higher than the previous five year mean (232 980, Figure 4.3.2.1). The median (3329) of the estimated 2SW returns in 2012 to Newfoundland was $9 \%$ lower than the previous year and $21 \%$ lower than the previous five year mean (4197, Figure 4.3.2.3).

## Quebec

The median of the estimated returns of small salmon in 2012 to Quebec ( 25 240) was 34\% lower than the previous year and $15 \%$ lower than the previous five year mean ( 29776 , Figure 4.3.2.1). The median of the estimated returns of 2 SW in 2012 to Quebec (27040)
was $28 \%$ lower than the previous year and $8 \%$ lower than the previous five year mean (29 256, Figure 4.3.2.3).

## Gulf of St Lawrence

The median of the estimated returns of small salmon in 2012 to the Gulf (17670) was $77 \%$ lower than the previous year and $68 \%$ lower than the previous five year mean ( 55970 , Figure 4.3.2.1). The median of the estimate of 2SW returns in 2012 to the Gulf (19 260) was $67 \%$ lower than the previous year and $34 \%$ lower than the previous five year mean (29 096, Figure 4.3.2.3).

## Scotia-Fundy

The median of the estimated returns of small salmon in 2012 to Scotia-Fundy (609) was $94 \%$ lower than the previous year and $94 \%$ lower than the previous five year mean (10 334, Figure 4.3.2.1). The median of the estimated 2SW returns in 2012 to Scotia-Fundy (1082) was $77 \%$ lower than the previous year and $61 \%$ lower than the previous five year mean (2753, Figure 4.3.2.3).

The model currently being used to extrapolate for the Nova Scotia Atlantic coast assessed rivers to total abundance (both returns and spawners) within SFAs 19-21 is likely leading to an overestimation of this portion of the regional abundance. The model is based on the assumption that the LaHave River salmon count is a representative index of this portion, an assumption that is likely invalid (ICES, 2010b). This issue only affects estimates since the closure of the recreational fisheries in the mid-2000s, and is expected to have very little effect on the advice provided on overall status of salmon in North America, but does have implications for regional management.

## USA

The estimated returns of small salmon in 2012 to USA (24) were $98 \%$ lower than the previous year and $96 \%$ lower than the previous five year mean (591, Figure 4.3.2.1). The estimated returns of 2SW in 2012 to USA (881) were $71 \%$ lower than the previous year and $51 \%$ lower than the previous five year mean (1782, Figure 4.3.2.3).

### 4.3.3 Estimates of spawning escapements

Updated estimates for small, large and 2SW spawners (1971 to 2012) were derived for the six geographic regions. A comparison between the numbers of small and large returns and spawners is presented in Figures 4.3.2.1 and 4.3.2.2. A comparison between the numbers of 2 SW returns, spawners, and CLs is presented in Figure 4.3.2.3.

## Canada

## Labrador

The median of the estimated numbers of 2SW spawners (21970) was $22 \%$ lower than the previous year and $18 \%$ higher than the previous five year mean ( 18642 ). The 2012 2SW spawners achieved $63 \%$ of the 2SW CL for Labrador (Figure 4.3.2.3). The 2SW CL has not been met during the time-series. The median of the estimated numbers of small spawners (170 700) was $37 \%$ lower than the previous year and $1 \%$ higher than the previous five year mean (168 228, Figure 4.3.2.1).

## Newfoundland

Finalized angling information from 2011 was used to update estimates of salmon spawners in that year. The median of the estimated numbers of 2SW spawners in 2012 (3304) was $9 \%$ lower than the previous year and $20 \%$ lower than the previous five year mean (4125). The 2012 2SW spawners achieved $82 \%$ of the 2 SW CL for Newfoundland. The 2SW CL has been met or exceeded in five out of the last ten years (Figure 4.3.2.3). The median of the estimated number of small spawners (213 800) was equal to the previous year and $4 \%$ higher than the previous five year mean (206 340, Figure 4.3.2.1). There was a general increase in both 2SW and 1SW spawners during the period 1992 to 1996 and 1998 to 2000, which is consistent with the closure of the commercial fisheries in Newfoundland.

## Quebec

The median of the estimated numbers of 2 SW spawners (20740) was $29 \%$ lower than the previous year and $8 \%$ lower than the previous five year mean (22 544). The 2012 2SW spawners achieved $68 \%$ of the 2 SW CL for Quebec (Figure 4.3.2.3). The median of the estimated number of small spawners (18410) was $34 \%$ lower than the previous year and $15 \%$ lower than the previous five year mean (21 590, Figure 4.3.2.1).

## Gulf of St Lawrence

The median of the estimated numbers of 2 SW spawners ( 18570 ) was $67 \%$ lower than the previous year and $34 \%$ lower than the previous five year mean (28 112). The 2012 2SW spawners achieved $63 \%$ of the 2SW CL for the Gulf (Figure 4.3.2.3). The 2SW CL has been met or exceeded in only one (2011) of the last ten years. The median of the estimated number of small spawners (10650) was $79 \%$ lower than the previous year and $71 \%$ lower than the previous five year mean (36 186, Figure 4.3.2.1).

## Scotia-Fundy

The median of the estimated numbers of 2SW spawners (1030) was $77 \%$ lower than the previous year and $61 \%$ lower than the previous five year mean (2645). The 2012 2SW spawners achieved $4 \%$ of the 2SW CL for Scotia-Fundy (Figure 4.3.2.3). The median of the estimated number of small spawners (590) was $94 \%$ lower than the previous year and $94 \%$ lower than the previous five year mean (10 174, Figure 4.3.2.1). As was the case with returns, these values may be overestimates (see Section 4.3.2).

## USA

The estimated numbers of 2SW spawners (2056) was $47 \%$ lower than the previous year and $12 \%$ lower than the previous five year mean (2329). The 2012 2SW spawners achieved $7 \%$ of the 2 SW CL for USA (Figure 4.3.2.3). The estimated number of small spawners (24) was $98 \%$ lower than the previous year and $96 \%$ lower than the previous five year mean (591, Figure 4.3.2.1).

### 4.3.4 Egg depositions in 2012

Egg depositions by all sea ages combined in 2012 exceeded or equalled the river-specific CLs in 31 of the 74 assessed rivers (42\%) and were less than $50 \%$ of CLs in 21 rivers ( $28 \%$ ) (Figure 4.3.4.1).

- One of the three (33\%) assessed rivers in Labrador exceeded its CL.
- In Newfoundland, $43 \%$ (six of 14) of assessed rivers exceeded their CLs. Three rivers (upper Exploits River, Rocky River and Little River) were below 50\% of their CLs.
- Two of the three ( $67 \%$ ) assessed rivers in the Gulf exceeded their CLs.
- In Quebec, $53 \%$ ( 21 of 40 ) of assessed rivers exceeded their CLs. Seven rivers were below $50 \%$ of their CLs.
- One of the seven ( $14 \%$ ) assessed rivers (North River) in Scotia-Fundy exceeded its CL. Four rivers were below $50 \%$ of CLs.
- Large deficiencies in egg depositions were noted in the USA. All seven assessed rivers were below $15 \%$ of their CLs.


### 4.3.5 Marine survival (return rates)

In 2012, return rate data were available from nine wild and four hatchery populations from rivers distributed among Newfoundland, Quebec, Scotia-Fundy, and USA (Tables 4.3.5.1 to 4.3.5.4). Wild return rates to 1 SW fish in 2012 decreased (range $10 \%$ to $96 \%$ ) relative to 2011 for eight of the nine assessed populations and increased ( $44 \%$ ) for one population (Conne River, Newfoundland). Large ( $>85 \%$ ) decreases were noted in 1SW return rates for the hatchery populations from 2011 to 2012.

Return rates in 2012 for wild 2SW salmon from the 2010 smolt class decreased relative to the 2009 smolt class on all the rivers (range $18 \%$ to $99 \%$ ). Similarly, return rates for 2 SW salmon decreased for the hatchery populations (range $35 \%$ to $94 \%$ ).

Analyses of time-series of return rates of smolts to 1 SW and 2SW adults by area (Tables 4.3.5.1 to 4.3.5.4; Figure 4.3.5.1) and analysis of the rates of change for individual rivers (Figure 4.3.5.2) provide insights into spatial and long and short-term temporal changes in marine survival of wild and hatchery populations.

- Return rates of wild populations exceed those of hatchery populations.
- 1 SW and 2SW return rates in 2012 decreased relative to 2011 for all areas for both wild and hatchery populations.
- Five year average return rates for 1SW wild salmon smolts returning to rivers of Newfoundland (Nfld) in 2008 to 2012 were similar to the averages for the previous period (2003 to 2007) for three monitored rivers and increased ( $25 \%$ and $70 \%$ ) on the other two.
- Five year average return rates for 1SW wild and hatchery salmon smolts returning to rivers of eastern North America (excluding Nfld) in 2008 to 2012 increased (range 8\% to $313 \%$ ) compared to the previous period (2003 to 2007) for all rivers monitored.
- Five year average return rates for 2SW wild and hatchery salmon smolts returning to rivers of eastern North America (excluding Nfld) in 2008 to 2012 increased (range $10 \%$ to $72 \%$ ) compared to the previous period (2003 to 2007) for six of the seven MSW rivers monitored.

Trends based on standardized return rates from the period 1970 to 2012 (Figure 4.3.5.1) include [N.B.no trend data are provided for Gulf, since there were no return rate data available for 2012]:

- 1 SW return rates of wild smolts to insular Newfoundland vary annually and have no temporal trend over the period 1970 to 2012 ( $\mathrm{p}>0.05$ ).
- 1 SW and 2 SW return rates of wild smolts to Quebec, although varying annually, have declined over the period 1983/84 to 2012 ( $\mathrm{p}<0.05$ ).
- 1 SW and 2 SW return rates of wild smolts to the Scotia-Fundy and USA, although varying annually, have no significant temporal trend over the period 1970 to 2012 ( $\mathrm{p}>0.05$ ).
- In Scotia Fundy and USA, hatchery smolt return rates to 2 SW salmon have decreased over the period 1970 to 2012 ( $\mathrm{p}<0.05$ ). 1SW return rates for Scotia Fundy stocks also declined for the period ( $\mathrm{p}<0.05$ ), while for USA there has been no significant trend ( $\mathrm{p}>0.05$ ).

Spatial trends include:

- 1 SW return rates for Newfoundland populations ( $3 \%$ to $10 \%$ ) in 2012 were greater than those for other populations in eastern North America ( $0.2 \%$ to $0.6 \%$ ).
- 1SW return rates in MSW salmon populations were greater than those of 2SW salmon within a smolt cohort for all monitored populations with the exception of one Quebec population (St Jean).


### 4.3.6 Pre-fisheries abundance

### 4.3.6.1 North American run-reconstruction model

The run-reconstruction model developed by Rago et al. (1993) and described in previous Working Group reports (ICES, 2008a; 2009a) and in the primary literature (Chaput et al., 2005) was used to estimate returns and spawners by size (small salmon, large salmon) and sea age group ( 2 SW salmon) to the six geographic regions of NAC. The input data were similar in structure to the data used previously by the Working Group (ICES, 2009a). Following on the recommendations from ICES (2008a), the run-reconstruction model for 2009 was developed using Monte Carlo simulation (OpenBUGS) similar to the approach applied for the NEAC area (Section 3.3.1). Estimates of returns and spawners to regions were provided for the time-series to 2012. The full set of data inputs and the summary output tables of catches, returns and spawners by sea age or size group are provided in Annex 6.

### 4.3.6.2 Non-maturing 1 SW salmon

The non-maturing component of 1SW fish, destined to be 2SW returns (excluding 3SW and previous spawners) is represented by the pre-fishery abundance estimator for year i designated as PFANAC1SW. This annual pre-fishery abundance is the estimated number of salmon in the North Atlantic on August 1st of the second summer at sea. As the prefishery abundance estimates for potential 2SW salmon requires estimates of returns to rivers, the most recent year for which an estimate of PFA is available is 2011. This is be-
cause pre-fishery abundance estimates for 2012 require 2SW returns to rivers in North America in 2013.

The medians derived from Monte Carlo simulations for 2 SW salmon returns by region and for NAC overall are shown in Figure 4.3.2.3. The estimated abundance of 2SW returns to rivers for NAC in 2012 was about 73680 fish ( $95 \%$ C.I. range 57280 to 90 500). The median estimate for 2012 is $45 \%$ lower than the previous year and $14 \%$ lower than the previous five year average ( 85942 ). The 2012 estimate ranks 35th (descending) out of the 43 year time-series.

The PFA estimates accounting for returns to rivers, fisheries at sea in North America, fisheries at West Greenland, and corrected for natural mortality are shown in Figure 4.3.6.2.1. The median of the estimates of non-maturing 1SW salmon in 2011 was 116800 fish ( $95 \%$ C.I. range 91770 to 144800 ). This value is $43 \%$ lower than the previous year (205 800) and $13 \%$ lower than the previous five year average (134 140). The estimated non-maturing 1SW salmon in 2011 ranks 30th (descending) out of the 41 year time-series.

### 4.3.6.3 Maturing 1 SW salmon

Maturing 1SW salmon are in some areas (particularly Newfoundland) a major component of salmon stocks, and their abundance when combined with that of the 2SW age group provides an index of the majority of an entire smolt cohort.
The medians of the region-specific estimates of returns of the 1 SW maturing component to rivers of NAC are summarized in Figure 4.3.2.1. Estimated abundance of 1SW returns in 2012 (458 500) was $28 \%$ lower than the previous year's estimate ( 640000 ) and $8 \%$ lower than the previous five year mean (500 020). While there was no estimated change to the levels of 1SW returns to Newfoundland in 2012 relative to 2011, decreases in 2012 were realized in Labrador (36\%), Quebec (34\%), Gulf (77\%), Scotia Fundy (94\%) and USA $(98 \%)$ relative to 2011. Returns of maturing 1SW salmon have generally increased over the time-series for the NAC, mainly as a result of the commercial fishery closures in Canada and increased returns over time to Labrador and Newfoundland.

The reconstructed distribution of the PFA of the 1SW maturing cohort of North American origin is shown in Figure 4.3.6.2.1. The estimated PFA of the maturing component in 2012 was 483000 fish, $28 \%$ lower than the previous year and $8 \%$ lower than the previous five year average (525780). Maximum abundance of the maturing cohort was estimated at over 910000 fish in 1981 and the recent estimate ranks 29th (descending) out of the 42 year time-series.

### 4.3.6.4 Total 1 SW recruits (maturing and non-maturing)

The pre-fishery abundance of 1SW maturing salmon and 1SW non-maturing salmon from North America from 1971-2011 (2012 PFA requires 2SW returns in 2013) were combined to give total recruits of 1 SW salmon (Figure 4.3.6.2.1). The PFA of the 1SW cohort, estimated for 2011, was 788800 fish, $11 \%$ higher than the 2010 PFA value ( 711000 ), $24 \%$ higher than the previous five year average (635 500). The 2011 PFA estimate ranks 22nd (descending) of the 41 year time-series. The abundance of the 1SW cohort has declined by $68 \%$ over the time-series from a peak of 1705000 fish in 1975.

### 4.3.7 Summary on status of stocks

In 2012, the midpoints of the spawner abundance estimates for all regions of NAC were below the CLs for 2SW salmon (Figure 4.3.2.3).

Estimates of PFA suggest continued low abundance of North American adult salmon. The total population of 1SW and 2SW Atlantic salmon in the Northwest Atlantic has oscillated around a generally declining trend since the 1970s with a period of persistent low abundance since the early 1990s. During 1993 to 2008, the total population of 1SW and 2SW Atlantic salmon was about 600000 fish, about half of the average abundance during 1972 to 1990. The maturing 1SW salmon in 2012 declined by $28 \%$ relative to 2011 and within the range of values for this age group over the period 1990 to 2012 (Figure 4.3.6.2.1). The non-maturing 1SW PFA for 2011 decreased by $43 \%$ from 2010 and remains among the lowest in the time-series.

Large declines of $1 \mathrm{SW}, 2 \mathrm{SW}$, and large salmon returns were noted in almost all areas in 2012 relative to 2011. The returns of 2SW fish in 2012 decreased from 2011 in all six geographic areas. Large declines (range $67 \%$ to $77 \%$, and levels among the lowest of the timeseries to 1971) were estimated for the three southern areas (USA, Scotia-Fundy, Gulf) and declines of $28 \%, 22 \%$ and $9 \%$ were estimated for Quebec, Labrador, and Newfoundland, respectively. Returns in 2012 of 1SW salmon relative to 2011 also decreased in all areas except for Newfoundland, and very large declines (range $77 \%$ to $98 \%$ ) were estimated along a north to south latitude gradient from Gulf to Scotia-Fundy to USA. Returns of 1SW salmon to Newfoundland in 2012 were similar to 2011.

The rank of the estimated returns in the 1971 to 2012 time-series and the proportions of the 2SW CLs achieved in 2012 for six regions in North America are shown below:

| ReGion | RANK OF 2012 RETURNS IN 1971 то 2012, (42 = LOWEST) |  | RANK OF 2003 то $(10=\mathrm{LOV}$ | ETURNS IN | Median estimate of 2SW SPAWNERS AS PERCENTAGE OF Conservation Limit |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | 2SW | 1SW | 2SW | (\%) |
| Labrador | 7 | 4 | 6 | 3 | 63 |
| Newfoundland | 7 | 31 | 5 | 9 | 82 |
| Quebec | 29 | 39 | 7 | 8 | 68 |
| Gulf | 42 | 36 | 10 | 9 | 63 |
| Scotia-Fundy | 42 | 41 | 10 | 10 | 4 |
| USA | 40 | 38 | 10 | 10 | 7 |

Egg depositions by all sea ages combined in 2012 exceeded or equalled the river-specific CLs in 45 of the 74 assessed rivers ( $61 \%$ ) and were less than $50 \%$ of CLs in 15 other rivers (20\%, Figure 4.3.4.1).

Despite major changes in fisheries, returns to southern regions (Scotia-Fundy and USA) have remained near historical lows and many populations are currently threatened with extirpation. In 2012, large declines in abundance from the higher abundances noted in 2011, were noted, reflecting an important mortality at sea on 1SW and 2SW salmon. The estimated PFA of 1SW non-maturing salmon ranked 30th (descending) of the 41-year time-series and the estimated PFA of 1SW maturing salmon ranked 29th (descending) of
the 42-year time-series. The continued low abundance of salmon stocks across North America, despite significant fishery reductions, and generally sustained smolt production (from the limited number of monitored rivers) strengthens the conclusions that factors acting on survival in the first and second years at sea are constraining abundance of Atlantic salmon.

Table 4.1.2.1. The number of professional and recreational gillnet licences issued at Saint-Pierre et Miquelon and reported landings.

| Number of licences |  | Reported Landings (tonnes) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Professional | Recreational | Professional | Recreational | Total |
| 1990 |  |  | 1.146 | 0.734 | 1.880 |
| 1991 |  |  | 0.632 | 0.530 | 1.162 |
| 1992 |  |  | 1.295 | 1.024 | 2.319 |
| 1993 |  |  | 1.902 | 1.041 | 2.943 |
| 1994 |  |  | 2.633 | 0.790 | 3.423 |
| 1995 | 12 | 42 | 0.392 | 0.445 | 0.837 |
| 1996 | 12 | 42 | 0.951 | 0.617 | 1.568 |
| 1997 | 6 | 36 | 0.762 | 0.729 | 1.491 |
| 1998 | 9 | 42 | 1.039 | 1.268 | 2.307 |
| 1999 | 7 | 40 | 1.182 | 1.140 | 2.322 |
| 2000 | 8 | 35 | 1.134 | 1.133 | 2.267 |
| 2001 | 10 | 42 | 1.544 | 0.611 | 2.155 |
| 2002 | 12 | 42 | 1.223 | 0.729 | 1.952 |
| 2003 | 12 | 42 | 1.620 | 1.272 | 2.892 |
| 2004 | 13 | 42 | 1.499 | 1.285 | 2.784 |
| 2005 | 14 | 52 | 2.243 | 1.044 | 3.287 |
| 2006 | 14 | 48 | 1.730 | 1.825 | 3.555 |
| 2007 | 13 | 53 | 0.970 | 0.977 | 1.947 |
| 2008 | 9 | 55 | Na | Na | 3.54 |
| 2009 | 8 | 50 | 1.87 | 1.59 | 3.46 |
| 2010 | 9 | 57 | 1.00 | 1.78 | 2.78 |
| 2011 | 9 | 56 | 1.76 | 1.99 | 3.75 |
| 2012 | 9 | 60 | 0.28 | 1.17 | 1.45 |

Table 4.1.3.1. Harvests (by weight) and the percent large by weight and number in the Aboriginal Peoples' Food, Social, and Ceremonial (FSC) fisheries in Canada.

## Aboriginal Peoples' FSC fisheries

| Year | Harvest $(\mathrm{t})$ | $\%$ large |  |
| :---: | :---: | :---: | :---: |
|  |  | by weight | by number |
| 1990 | 31.9 | 78 |  |
| 1991 | 29.1 | 87 |  |
| 1992 | 34.2 | 83 | 58 |
| 1993 | 42.6 | 83 | 56 |
| 1994 | 41.7 | 83 | 65 |
| 1995 | 32.8 | 82 | 74 |
| 1996 | 47.9 | 87 | 63 |
| 1997 | 39.4 | 91 | 49 |
| 1998 | 47.9 | 83 | 41 |
| 1999 | 45.9 | 73 | 47 |
| 2000 | 45.7 | 68 | 43 |
| 2001 | 42.1 | 72 | 49 |
| 2002 | 46.3 | 68 | 44 |
| 2003 | 44.3 | 72 | 34 |
| 2004 | 60.8 | 66 | 39 |
| 2005 | 56.7 | 57 | 40 |
| 2006 | 61.4 | 60 | 44 |
| 2007 | 48.0 | 62 | 45 |
| 2008 | 62.4 | 66 | 38 |
| 2009 | 51.1 | 65 | 41 |
| 2010 | 59.3 | 59 | 44 |
| 2011 | 70.4 | 63 |  |
| 2012 |  |  | 4 |

Table 4.1.3.2. Harvests (by weight) and the percent large by weight and number in the Resident Food Fishery in Labrador, Canada.

| Labrador resident food fishery |  |  |  |
| :--- | :---: | :---: | :---: |
| Year | Harvest (t) | by weight | \% large |
|  |  | 30 | by number |
| 2000 | 3.5 | 33 | 18 |
| 2001 | 4.6 | 27 | 23 |
| 2002 | 6.1 | 32 | 15 |
| 2003 | 6.7 | 40 | 21 |
| 2004 | 2.2 | 32 | 26 |
| 2005 | 2.7 | 39 | 20 |
| 2006 | 2.6 | 23 | 27 |
| 2007 | 1.7 | 46 | 13 |
| 2008 | 2.3 | 42 | 25 |
| 2009 | 2.9 | 38 | 28 |
| 2010 | 2.3 | 51 | 26 |
| 2011 | 2.1 | 48 | 37 |
| 2012 | 1.7 |  | 32 |

Table 4.1.3.3. Harvests of small and large salmon, and the percent large by number, in the recreational fisheries of Canada, 1974 to 2012. The values for 2012 are provisional.

| Year | Small | Large | Bоth size Groups | \% LarGe |
| :---: | :---: | :---: | :---: | :---: |
| 1974 | 53887 | 31720 | 85607 | 37\% |
| 1975 | 50463 | 22714 | 73177 | 31\% |
| 1976 | 66478 | 27686 | 94164 | 29\% |
| 1977 | 61727 | 45495 | 107222 | 42\% |
| 1978 | 45240 | 28138 | 73378 | 38\% |
| 1979 | 60105 | 13826 | 73931 | 19\% |
| 1980 | 67314 | 36943 | 104257 | 35\% |
| 1981 | 84177 | 24204 | 108381 | 22\% |
| 1982 | 72893 | 24640 | 97533 | 25\% |
| 1983 | 53385 | 15950 | 69335 | 23\% |
| 1984 | 66676 | 9982 | 76658 | 13\% |
| 1985 | 72389 | 10084 | 82473 | 12\% |
| 1986 | 94046 | 11797 | 105843 | 11\% |
| 1987 | 66475 | 10069 | 76544 | 13\% |
| 1988 | 91897 | 13295 | 105192 | 13\% |
| 1989 | 65466 | 11196 | 76662 | 15\% |
| 1990 | 74541 | 12788 | 87329 | 15\% |
| 1991 | 46410 | 11219 | 57629 | 19\% |
| 1992 | 77577 | 12826 | 90403 | 14\% |
| 1993 | 68282 | 9919 | 78201 | 13\% |
| 1994 | 60118 | 11198 | 71316 | 16\% |
| 1995 | 46273 | 8295 | 54568 | 15\% |
| 1996 | 66104 | 9513 | 75617 | 13\% |
| 1997 | 42891 | 6756 | 49647 | 14\% |
| 1998 | 45810 | 4717 | 50527 | 9\% |
| 1999 | 43667 | 4811 | 48478 | 10\% |
| 2000 | 45811 | 4627 | 50438 | 9\% |
| 2001 | 43353 | 5571 | 48924 | 11\% |
| 2002 | 43904 | 2627 | 46531 | 6\% |
| 2003 | 38367 | 4694 | 43061 | 11\% |
| 2004 | 43124 | 4578 | 47702 | 10\% |
| 2005 | 33922 | 4132 | 38054 | 11\% |
| 2006 | 33668 | 3014 | 36682 | 8\% |
| 2007 | 26279 | 3499 | 29778 | 12\% |
| 2008 | 46458 | 2839 | 49297 | 6\% |
| 2009 | 32944 | 3373 | 36317 | 9\% |
| 2010 | 45407 | 3209 | 48616 | 7\% |
| 2011 | 49931 | 4141 | 54072 | 8\% |
| 2012 | 35040 | 2680 | 37720 | 7\% |

Table 4.1.3.4. Numbers of salmon hooked and-released in Eastern Canadian salmon angling fisheries. Blank cells indicate no data. Released fish in the kelt fishery of New Brunswick are not included in the totals for New Brunswick nor Canada. Totals for all years prior to 1997 are incomplete and are considered minimal estimates.

|  | Newfoundland |  |  | Nova Scotia |  |  | New Brunswick |  |  |  |  | Prince Edward Island |  |  | Quebec |  |  | CANADA |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Small | Large | Total | Small | Large | Total | Small Kelt | Small Bright | Large Kelt | Large Bright | Total | Small | Large | Total | Small | Large | Total | SMALL | LARGE | TOTAL |
| 1984 |  |  |  | 939 | 1,655 | 2,594 | 661 | 851 | 1,020 | 14,479 | 15,330 |  |  |  |  |  |  | 1,790 | 16,134 | 17,924 |
| 1985 |  | 315 | 315 | 1,323 | 6,346 | 7,669 | 1,098 | 3,963 | 3,809 | 17,815 | 21,778 |  |  | 67 |  |  |  | 5,286 | 24,476 | 29,762 |
| 1986 |  | 798 | 798 | 1,463 | 10,750 | 12,213 | 5,217 | 9,333 | 6,941 | 25,316 | 34,649 |  |  |  |  |  |  | 10,796 | 36,864 | 47,660 |
| 1987 |  | 410 | 410 | 1,311 | 6,339 | 7,650 | 7,269 | 10,597 | 5,723 | 20,295 | 30,892 |  |  |  |  |  |  | 11,908 | 27,044 | 38,952 |
| 1988 |  | 600 | 600 | 1,146 | 6,795 | 7,941 | 6,703 | 10,503 | 7,182 | 19,442 | 29,945 | 767 | 256 | 1,023 |  |  |  | 12,416 | 27,093 | 39,509 |
| 1989 |  | 183 | 183 | 1,562 | 6,960 | 8,522 | 9,566 | 8,518 | 7,756 | 22,127 | 30,645 |  |  |  |  |  |  | 10,080 | 29,270 | 39,350 |
| 1990 |  | 503 | 503 | 1,782 | 5,504 | 7,286 | 4,435 | 7,346 | 6,067 | 16,231 | 23,577 |  |  | 1,066 |  |  |  | 9,128 | 22,238 | 31,366 |
| 1991 |  | 336 | 336 | 908 | 5,482 | 6,390 | 3,161 | 3,501 | 3,169 | 10,650 | 14,151 | 1,103 | 187 | 1,290 |  |  |  | 5,512 | 16,655 | 22,167 |
| 1992 | 5,893 | 1,423 | 7,316 | 737 | 5,093 | 5,830 | 2,966 | 8,349 | 5,681 | 16,308 | 24,657 |  |  | 1,250 |  |  |  | 14,979 | 22,824 | 37,803 |
| 1993 | 18,196 | 1,731 | 19,927 | 1,076 | 3,998 | 5,074 | 4,422 | 7,276 | 4,624 | 12,526 | 19,802 |  |  |  |  |  |  | 26,548 | 18,255 | 44,803 |
| 1994 | 24,442 | 5,032 | 29,474 | 796 | 2,894 | 3,690 | 4,153 | 7,443 | 4,790 | 11,556 | 18,999 | 577 | 147 | 724 |  |  |  | 33,258 | 19,629 | 52,887 |
| 1995 | 26,273 | 5,166 | 31,439 | 979 | 2,861 | 3,840 | 770 | 4,260 | 880 | 5,220 | 9,480 | 209 | 139 | 348 |  | 922 | 922 | 31,721 | 14,308 | 46,029 |
| 1996 | 34,342 | 6,209 | 40,551 | 3,526 | 5,661 | 9,187 |  |  |  |  |  | 472 | 238 | 710 |  | 1,718 | 1,718 | 38,340 | 13,826 | 52,166 |
| 1997 | 25,316 | 4,720 | 30,036 | 713 | 3,363 | 4,076 | 3,457 | 4,870 | 3,786 | 8,874 | 13,744 | 210 | 118 | 328 | 182 | 1,643 | 1,825 | 31,291 | 18,718 | 50,009 |
| 1998 | 31,368 | 4,375 | 35,743 | 688 | 2,476 | 3,164 | 3,154 | 5,760 | 3,452 | 8,298 | 14,058 | 233 | 114 | 347 | 297 | 2,680 | 2,977 | 38,346 | 17,943 | 56,289 |
| 1999 | 24,567 | 4,153 | 28,720 | 562 | 2,186 | 2,748 | 3,155 | 5,631 | 3,456 | 8,281 | 13,912 | 192 | 157 | 349 | 298 | 2,693 | 2,991 | 31,250 | 17,470 | 48,720 |
| 2000 | 29,705 | 6,479 | 36,184 | 407 | 1,303 | 1,710 | 3,154 | 6,689 | 3,455 | 8,690 | 15,379 | 101 | 46 | 147 | 445 | 4,008 | 4,453 | 37,347 | 20,526 | 64,482 |
| 2001 | 22,348 | 5,184 | 27,532 | 527 | 1,199 | 1,726 | 3,094 | 6,166 | 3,829 | 11,252 | 17,418 | 202 | 103 | 305 | 809 | 4,674 | 5,483 | 30,052 | 22,412 | 59,387 |
| 2002 | 23,071 | 3,992 | 27,063 | 829 | 1,100 | 1,929 | 1,034 | 7,351 | 2,190 | 5,349 | 12,700 | 207 | 31 | 238 | 852 | 4,918 | 5,770 | 32,310 | 15,390 | 50,924 |
| 2003 | 21,379 | 4,965 | 26,344 | 626 | 2,106 | 2,732 | 1,555 | 5,375 | 1,042 | 7,981 | 13,356 | 240 | 123 | 363 | 1,238 | 7,015 | 8,253 | 28,858 | 22,190 | 53,645 |
| 2004 | 23,430 | 5,168 | 28,598 | 828 | 2,339 | 3,167 | 1,050 | 7,517 | 4,935 | 8,100 | 15,617 | 135 | 68 | 203 | 1,291 | 7,455 | 8,746 | 33,201 | 23,130 | 62,316 |
| 2005 | 33,129 | 6,598 | 39,727 | 933 | 2,617 | 3,550 | 1,520 | 2,695 | 2,202 | 5,584 | 8,279 | 83 | 83 | 166 | 1,116 | 6,445 | 7,561 | 37,956 | 21,327 | 63,005 |
| 2006 | 30,491 | 5,694 | 36,185 | 1,014 | 2,408 | 3,422 | 1,071 | 4,186 | 2,638 | 5,538 | 9,724 | 128 | 42 | 170 | 1,091 | 6,185 | 7,276 | 36,910 | 19,867 | 60,486 |
| 2007 | 17,719 | 4,607 | 22,326 | 896 | 1,520 | 2,416 | 1,164 | 2,963 | 2,067 | 7,040 | 10,003 | 63 | 41 | 104 | 951 | 5,392 | 6,343 | 22,592 | 18,600 | 41,192 |
| 2008 | 25,226 | 5,007 | 30,233 | 1,016 | 2,061 | 3,077 | 1,146 | 6,361 | 1,971 | 6,130 | 12,491 | 3 | 9 | 12 | 1,361 | 7,713 | 9,074 | 33,967 | 20,920 | 54,887 |
| 2009 | 26,681 | 4,272 | 30,953 | 670 | 2,665 | 3,335 | 1,338 | 2,387 | 1,689 | 8,174 | 10,561 | 6 | 25 | 31 | 1,091 | 6,180 | 7,271 | 30,835 | 21,316 | 52,151 |
| 2010 | 27,256 | 5,458 | 32,714 | 717 | 1,966 | 2,683 | 463 | 5,730 | 1,920 | 5,660 | 11,390 | 42 | 27 | 69 | 1,356 | 7,683 | 9,039 | 35,101 | 20,794 | 55,895 |
| 2011 | 26,240 | 8,119 | 34,359 | 1,157 | 4,320 | 5,477 |  | 6,537 |  | 12,466 | 19,003 | 46 | 46 | 92 | 3,100 | 9,327 | 12,427 | 37,080 | 34,278 | 71,358 |
| 2012 | 27,629 | 5,759 | 33,388 | 272 | 1,017 | 1,289 |  | 2,504 |  | 5,330 | 7,834 |  |  | 0 | 2,126 | 6,174 | 8,300 | 32,531 | 18,280 | 50,811 |

Table 4.1.4.1. Reported harvests expressed as 2SW salmon equivalents in North American salmon fisheries. Only midpoints of the estimated values have been used.

| Year (i) | MIXED STOCK |  |  |  |  | CANADA |  |  |  |  |  |  | $\begin{array}{\|c} \begin{array}{c} \text { North } \\ \text { American } \\ \text { Total } \end{array} \\ \hline \end{array}$ | $\begin{aligned} & \text { Teminal } \\ & \text { Fishheres } \\ & \text { as a of } \\ & \text { N N ot otal } \end{aligned}$ | Greenland Total | $\begin{gathered} \text { NW } \\ \text { Atlantic } \\ \text { Total } \\ \hline \end{gathered}$ | Harvest in homewatersas $\%$ of total NW Atlantic | Estimatedabundance inNorth America | Exploitation rates in North America on 2SW equivalents |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | TERMINAL FISHERRIES IN (Yeari) |  |  |  |  |  |  |  |  |  |  |  |  |
|  | NF-LAB <br> Comm/ <br> Food 1SW <br> (Year i-1) | $\begin{gathered} \hline \begin{array}{c} \text { \% } 1 \text { SW of } \\ \text { total 2SW } \\ \text { equivalents } \\ (\text { Year i) } \end{array} \\ \hline \end{gathered}$ | NF-LAB Comm/ Food 2SW (Yeari) (a) | NF-Lab Comm/Food total (Yeari) |  | Labrador | Newfoundland | Quebec | Gulf | $\begin{aligned} & \text { Scotia - } \\ & \text { F Fundy } \end{aligned}$ | Canadian | USA |  |  |  |  |  |  |  |
| 1972 | 20122 | 0.12 | 153816 | 173938 |  |  | 594 | 27400 | 20260 | 5610 | 54294 |  | 228577 |  | 198063 | 426640 |  | 302400 |  |
| 1973 | 17448 | 0.07 | 21924 | 236671 | 0 | 1010 | 770 | 32730 | 15460 | 6211 | 56181 | 327 | 293179 | 19 | 148170 | 441350 | 66 | 377000 | 0.78 |
| 1974 | 23731 | 0.09 | 235915 | 259646 |  | 810 | 496 | 47540 | 18990 | 13040 | 79976 | 247 | 339869 | 24 | 186561 | 526430 | 65 | 449500 |  |
| 1975 | 23467 | 9 | 237565 | 261032 | 0 | 330 | 504 | 41080 | 14550 | 12510 | 68474 | 389 | 329895 | 21 | 154569 | 48463 | 68 | 416500 | 0.79 |
| 1976 | 35044 | 0.12 | 256586 | 291630 | 323 | 830 | 378 | 42110 | 16160 | 11120 | 70598 | 191 | 362742 | 20 | 194397 | 557139 | 65 | 431500 | 0.84 |
| 1977 | 26757 | 0.10 | 241253 | 268010 |  | 1280 | 781 | 42270 | 29170 | 13440 | 86941 | 1355 | 356306 | 25 | 11281 | 469177 | 76 | 47340 | 0.75 |
| 1978 | 26994 | 0.15 | 157309 | 184303 | 0 |  | 536 | 37430 | 20340 | 9361 | 68437 | 894 | 253634 | 27 | 142706 | 396341 | 64 | 317400 | 0.80 |
| 1979 | 13501 | 0.13 | 92066 | 105567 | 0 | 609 | 126 | 25240 | 623 | 3831 | 36059 | 433 | 142059 | 26 | 103741 | 245799 | 58 | 172100 | 0.83 |
| 1980 | 20623 | 0.09 | 217186 | 237809 | 0 | 890 | 641 | 53570 | 25350 | 17370 | 97821 | 1533 | 337163 | 29 | 14184 | 479007 | 70 | 451800 | 0.75 |
| 1981 | 33738 | 0.14 | 201367 | 235105 | 0 | 520 | 438 | 4400 | 14625 | 12850 | 72833 | 1267 | 309205 | 24 | 120923 | 430128 | 72 | 365500 | 0.85 |
| 1982 | 33596 | 0.20 | 134407 | 168002 |  | 620 | 396 | 35300 | 20780 | 8927 | 66023 | 1413 | 235438 | 29 | 16183 | 396621 | 59 | 291200 | 0.81 |
| 1983 | 25254 | 0.18 | 11601 | 136855 | 323 | 428 | 423 | 3440 | 17330 | 12270 | 64891 | 386 | 20245 | 32 | 145870 | 348325 | 58 | 237200 | 0.85 |
| 1984 | 19052 | 0.19 | 82808 | 101860 | 323 | 510 | 183 | 24820 | 3560 | 3960 | 33033 | 675 | 135891 | 25 | 26837 | 162729 | 84 | 204900 | 0.66 |
| 1985 | 14340 | 0.15 | 78761 | 93101 | 323 | 294 | 15 | 27820 | 990 | 5060 | 34179 | 645 | 128248 | 27 | 32431 | 16079 | 80 | 218100 | 0.59 |
| 1986 | 19601 | 0.16 | 104905 | 124506 | 269 | 467 | 34 | 34220 | 1650 | 2940 | 39311 | 606 | 164692 | 24 | 99140 | 263832 | 62 | 273300 | 0.60 |
| 1987 | 24801 | 0.16 | 132175 | 156975 | 215 | 640 |  | 34220 | 2000 | 1440 | 38316 | 300 | 195807 |  | 123511 | 319318 | 61 | 267200 | 0.73 |
| 1988 | 31578 | 0.28 | 81139 | 112717 | 215 | 710 | 18 | 34630 | 1370 | 1460 | 38188 | 248 | 151368 | 25 | 123799 | 275167 | 55 | 222300 | 0.68 |
| 1989 | 21910 | 0.21 | 81362 | 103272 | 215 | 461 |  | 29320 | 1230 | 330 | 31345 | 397 | 135229 |  | 84905 | 220134 | 61 | 20140 |  |
| 1990 | 19289 | 0.25 | 57363 | 76652 | 205 | 357 | 22 | 28430 | 1100 | 660 | 30569 | 695 | 108121 | 29 | 43646 | 151767 | 71 | 181500 | 0.60 |
| 1991 | 11842 | 0.23 | 40438 | 52280 | 129 |  | 12 | 29650 | 830 | 1380 | 31965 | 231 | 84605 |  | 5223 | 136828 | 62 | 154500 | 0.55 |
| 1992 | 9851 | 0.28 | 25105 | 34957 | 248 | 782 |  | 30480 | 1180 | 1150 | 33592 | 167 | 68963 | 49 | 79585 | 145548 | 46 | 152300 | 0.45 |
| 1993 | 3110 | 0.19 | 13276 | 16385 | 312 | 387 | 0 | 23540 | 560 | 1158 | 25645 | 166 | 42509 | 61 | 29807 | 72315 | 59 | 126800 | 0.34 |
| 1994 | 2077 | 0.15 | 11936 | 14014 | 366 | 490 | 0 | 24580 | 710 | 777 | 26557 | 2 | 40939 | 65 | 1890 | 42829 | 96 | 112300 | 0.36 |
| 1995 | 1183 | 0.12 | 8678 | 9861 |  | 460 | 0 | 23700 | 530 | 365 | 25055 | 0 | 35002 | 72 | 1890 | 36892 | 95 | 139900 | 0.25 |
| 1996 | 1033 | 0.15 | 5646 | 6679 | 172 | 380 | 0 | 22690 | 860 | 819 | 24749 | 0 | 31600 | 78 | 19181 | 50781 | 62 | 120200 | 0.26 |
| 1997 | 943 | 0.15 | 5391 | 6334 | 161 | 210 | 0 | 18620 | 820 | 600 | 20250 | 0 | 26745 | 76 | 19332 | 46077 | 58 | 98300 | 0.27 |
| 1998 | 1131 | 0.39 | 1761 | 2892 | 248 | 203 | 0 | 11280 | 490 | 331 | 12304 | 0 | 15444 | 80 | 13048 | 28492 | 54 | 68110 | 0.23 |
| 1999 | 175 | 0.17 | 842 | 1017 | 250 | 270 | 0 | 9190 | 790 | 460 | 10710 | 0 | 11977 | 89 | 4322 | 16299 | 73 | 71090 | 0.17 |
| 2000 | 151 | 0.13 | 1050 | 1201 | 244 | 260 | 0 | 8890 | 600 | 200 | 9950 | 0 | 11395 | 87 | 6444 | 17838 | 64 | 72820 | 0.16 |
| 2001 | 285 | 0.18 | 1336 | 1621 | 232 | 320 | 0 | 9660 | 930 | 264 | 11174 | 0 | 13027 | 86 | 5933 | 18960 | 69 | 83970 | 0.16 |
| 2002 | 261 | 0.20 | 1078 | ${ }^{1340}$ | 210 | 200 |  | 6200 | 540 | 183 | 7123 | 0 | 8673 | 82 | 8606 | 17279 | 50 | 53640 | 0.16 |
| 2003 | 310 | 0.16 | 1689 | 1998 | 311 | 233 | 0 | 8510 | 820 | 207 | 9770 | 0 | 12080 | 81 | 3224 | 15303 | 79 | 81320 | 0.15 |
| 2004 | 352 | 0.11 | 2870 | 3222 | 300 | 270 |  | 8410 | 850 | 116 | 9646 | 0 | 13167 | 73 | 3475 | 16642 | 79 | 79290 | 0.17 |
| 2005 | 465 | 0.18 | 2187 | 2652 | 354 | 280 | 0 | 7460 | 960 | 107 | 8807 | 0 | 11813 | 75 | 4338 | 16151 | 73 | 80690 | 0.15 |
| 2006 | 560 | 0.19 | 2399 | 2959 | 383 | 220 |  | 7140 | 820 | 151 | 8331 | 0 | 11673 | 71 | 4181 | 15853 | 74 | 76940 | 0.15 |
| 2007 | 560 | 0.21 | 2059 | 2619 | 210 | 230 |  | 670 | 860 | 109 | 7899 | 0 | 10728 | 74 | 4935 | 15662 | 68 | 72220 | 0.15 |
| 2008 | 496 | 0.14 | 3035 | 3531 | 381 | 230 | 0 | 6470 | 850 | 0 | 7550 | 0 | 11462 | 66 | 6616 | 18078 | 63 | 78870 | 0.15 |
| 2009 | 540 | 0.17 | 2596 | 3136 | 372 | 220 | $\bigcirc$ | 6510 | 890 | 0 | 7620 | 0 | 11129 | 68 | 7549 | 18678 | 60 | 92060 | 0.12 |
| 2010 | 441 | 0.13 | 2892 | 3333 | 299 | 198 | 0 | 5870 | 830 | 0 | 6898 | 0 | 10530 | 66 | 6671 | 17201 | 61 | 72840 | 0.14 |
| 2011 | 540 | 0.14 | 3453 | 3993 | 404 | 140 |  | 8010 | 1490 | 0 | 9640 | 0 | 14038 | 69 | 8764 | 22801 | 62 | 143000 | 0.10 |
| 2012 | 613 552 | 0.16 | 3284 | 3897 | 156 | 90 | $\bigcirc$ | 6300 | 690 | 0 | 7080 | 0 | 11133 | 64 | 6331 | 17464 | 64 | 79470 | 0.14 |

NF-Lab comm as $1 \mathrm{SW}=\mathrm{NCl}($ mid-pt) $* 0.677057$ (M of 0.03 per month for 13 months to July for Canadian terminal fisheries)
NF-Lab comm as $2 \mathrm{SW}=\mathrm{NC2}$ (mid-pt) $* 0.970446$ ( M of 0.03 per month for 1 month to July of Canadian terminal fisheries)
Te-starting in 1998 , there was no commercial fishery in Labrador; numbers reflect size of aboriginal fish harvest in 1998 -2012 and resident food fishery harvest in 2000 -2012

Table 4.3.1.1. Estimated smolt production by smolt migration year in monitored rivers of eastern North America, 1991 to 2012.

| Smolt <br> Migration <br> Year | USA <br> Narraguagus | Scotia-Fundy |  |  |  | Gulf |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Nashwaak | Big Salmon | LaHave | St. Mary's (West Br.) | Margaree | Northwest Miramichi | Southwest Miramichi | Restigouche | Kedgwick |
| 1991 |  |  |  |  |  |  |  |  |  |  |
| 1992 |  |  |  |  |  |  |  |  |  |  |
| 1993 |  |  |  |  |  |  |  |  |  |  |
| 1994 |  |  |  |  |  |  |  |  |  |  |
| 1995 |  |  |  |  |  |  |  |  |  |  |
| 1996 |  |  |  | 20510 |  |  |  |  |  |  |
| 1997 | 2898 |  |  | 16550 |  |  |  |  |  |  |
| 1998 | 2866 | 22750 |  | 15600 |  |  |  |  |  |  |
| 1999 | 4346 | 28500 |  | 10420 | - | - | 390500 |  |  |  |
| 2000 | 2094 | 15800 |  | 16300 |  |  | 162000 |  |  |  |
| 2001 | 2621 | 11000 | 5100 | 15700 | - |  | 220000 | 306300 |  |  |
| 2002 | 1800 | 15000 | 4300 | 11860 |  | 63200 | 241000 | 711400 |  |  |
| 2003 | 1368 | 9000 | 9200 | 14034 |  | 83100 | 286000 | 48500 | 379000 | 91800 |
| 2004 | 1344 | 13600 | 6000 | 21613 |  | 105800 | 368000 | 1167000 | 449000 | 131500 |
| 2005 | 1298 | 5200 | 4550 | 5270 | 7350 | 94200 | 151200 |  | 630000 | 67000 |
| 2006 | 2612 | 25400 |  | 22971 | 25100 | 113700 | 435000 | 1330000 | 500000 | 129000 |
| 2007 | 1240 | 21550 |  | 24430 | 16110 | 112400 |  | 1344000 | 1087000 | 116600 |
| 2008 | 1029 | 7310 |  | 14450 | 15217 | 128800 |  | 901500 | 486800 | 110100 |
| 2009 | 1180 | 15900 |  | 8643 | 14820 | 96800 |  | 1035000 | 491000 | 126800 |
| 2010 | 2170 | 12500 |  | 16215 |  |  |  | 2165000 | 636600 | 108600 |
| 2011 | 1404 | 8750 |  |  |  |  | 768000 |  | 792000 | 275178 |
| 2012 | 969 | 11060 |  |  |  |  |  |  | 842000 | 155012 |

Table 4.3.1.1 (continued). Estimated smolt production by smolt migration year in monitored rivers of eastern North America, 1991 to 2012.

| Smolt <br> Migration Year | Quebec |  | Newfoundland |  | NE Trepassey | Campbellton | Western Arm Brook | LABRADOR <br> Sand Hill River |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | St. Jean | De la Trinite | Conne | Rocky |  |  |  |  |
| 1991 | 113927 | 40863 | 74645 | 7732 | 1911 |  | 13453 |  |
| 1992 | 154980 | 50869 | 68208 | 7813 | 1674 |  | 15405 |  |
| 1993 | 142972 | 86226 | 55765 | 5115 | 1849 | 31577 | 13435 |  |
| 1994 | 74285 | 55913 | 60762 | 9781 | 944 | 41663 | 9283 |  |
| 1995 | 60227 | 71899 | 62749 | 7577 | 792 | 39715 | 15144 |  |
| 1996 | 104973 | 61092 | 94088 | 14261 | 1749 | 58369 | 14502 |  |
| 1997 |  | 31892 | 100983 | 16900 | 1829 | 62050 | 23845 |  |
| 1998 | 95843 | 28962 | 69841 | 12163 | 1727 | 50441 | 17139 |  |
| 1999 | 114255 | 56557 | 63658 | 8625 | 1419 | 47256 | 13500 |  |
| 2000 | 50993 | 39744 | 60777 | 7616 | 1740 | 35596 | 12706 |  |
| 2001 | 109845 | 70318 | 86899 | 9392 | 916 | 37170 | 16013 |  |
| 2002 | 71839 | 44264 | 81806 | 10144 | 2074 | 32573 | 14999 |  |
| 2003 | 60259 | 53030 | 71479 | 4440 | 1064 | 35089 | 12086 |  |
| 2004 | 54821 | 27051 | 79667 | 13047 | 1571 | 32780 | 17323 |  |
| 2005 | 96002 | 34867 | 66196 | 15847 | 1384 | 30123 | 8607 |  |
| 2006 | 102939 |  | 35487 | 13200 | 1385 | 33302 | 20826 |  |
| 2007 | 135360 | 42923 | 63738 | 12355 | 1777 | 35742 | 16621 |  |
| 2008 | 45978 | 35036 | 68242 | 18338 | 1868 | 40390 | 17444 |  |
| 2009 | 37297 | 32680 | 71085 | 14041 | 1600 | 36722 | 18492 |  |
| 2010 | 47187 | 37500 | 54392 | 15098 | 1012 | 41069 | 19044 |  |
| 2011 | 45050 | 44400 | 50701 | 9311 | 800 | 37033 | 20544 |  |
| 2012 | 40787 | 45108 | 51220 | 5673 | 1557 | 44193 | 13573 | 82537 |

Table 4.3.5.1. Return rates (\%), by year of smolt migration, of wild Atlantic salmon to 1SW (or small) salmon to North American rivers, 1991 to 2011 . The year 1991 was selected for illustration as it is the first year of the commercial fishery moratorium for the island of Newfoundland.

| Sмо <br> LT <br> YEAR | USA <br> Narragua gus | SF <br> Nashw aak | LaHa ve | St.Mar <br> y's | Gulf <br> Marga ree | NWMiram ichi | SW <br> Miramc ihi | Mirami chi | Quebec |  |  |  | Nfid |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  | à la <br> bar <br> be | Sai <br> nt <br> Jea <br> n | Be c sci e | de la <br> Trini te | Highla nds | Con ne | Roc ky | NE <br> Trepas sey | Campbell ton | WA B |
| 1991 |  |  |  |  |  |  |  |  | 0.6 | 0.5 | 1.2 | 1.6 |  | 3.4 | 3.1 | 2.6 |  | 3.6 |
| 1992 |  |  |  |  |  |  |  |  | 0.5 | 0.4 | 1.3 | 0.8 |  | 4.0 | 3.7 | 4.7 |  | 6.1 |
| 1993 |  |  |  |  |  |  |  |  | 0.4 | 0.3 | 0.9 | 0.7 | 1.5 | 2.7 | 3.1 | 5.4 | 9.0 | 7.1 |
| 1994 |  |  |  |  |  |  |  |  |  | 0.3 | 1.2 | 0.6 | 1.6 | 5.8 | 3.9 | 8.5 | 7.3 | 8.9 |
| 1995 |  |  |  |  |  |  |  |  |  | 0.6 | 1.4 | 0.9 | 1.6 | 7.2 | 4.7 | 9.2 | 8.1 | 8.1 |
| 1996 |  |  | 1.5 |  |  |  |  |  |  | 0.3 |  | 0.6 | 3.2 | 3.4 | 3.1 | 2.9 | 3.4 | 3.5 |
| 1997 | 0.04 |  | 4.3 |  |  |  |  |  |  |  |  | 1.7 | 1.4 | 2.9 | 2.5 | 5.0 | 5.3 | 7.2 |
| 1998 | 0.22 | 2.9 | 2.0 |  |  |  |  |  |  | 0.3 |  | 1.4 | 2.5 | 3.4 | 2.7 | 4.9 | 6.1 | 6.1 |
| 1999 | 0.30 | 1.8 | 4.8 |  |  | 3.0 |  |  |  | 0.3 |  | 0.4 | 0.6 | 8.1 | 3.2 | 5.9 | 3.8 | 11.1 |
| 2000 | 0.25 | 1.5 | 1.2 |  |  | 4.9 |  |  |  | 0.5 |  | 0.3 | 0.6 | 2.5 | 3.1 | 3.2 | 6.0 | 4.4 |
| 2001 | 0.16 | 3.1 | 2.7 |  |  | 6.6 | 8.6 | 7.9 |  | 0.5 |  | 0.6 |  | 3.0 | 2.9 | 7.1 | 5.3 | 9.2 |
| 2002 | 0.00 | 1.9 | 2.0 |  | 1.5 | 2.4 | 3.0 | 3.0 |  | 0.6 |  | 0.9 |  | 2.4 | 4.0 | 5.5 | 6.8 | 9.4 |
| 2003 | 0.08 | 6.4 | 1.8 |  | 1.6 | 4.1 | 6.8 | 5.9 |  | 0.6 |  | 0.6 |  | 5.3 | 3.8 | 6.6 | 7.8 | 9.5 |
| 2004 | 0.08 | 5.1 | 1.1 |  | 0.9 | 2.6 | 1.8 | 2.0 |  | 0.7 |  | 1.0 |  | 2.5 | 3.3 | 4.4 | 11.4 | 5.9 |
| 2005 | 0.24 | 12.7 | 8.0 | 3.1 | 1.1 | 3.6 |  |  |  | 0.4 |  | 1.5 |  | 4.0 | 2.2 | 5.5 | 9.2 | 15.1 |
| 2006 | 0.09 | 1.8 | 1.5 | 0.7 | 0.7 | 1.4 | 1.5 | 1.5 |  | 0.3 |  |  |  | 3.3 | 1.3 | 2.7 | 5.6 | 3.8 |
| 2007 | 0.33 | 5.6 | 2.3 | 1.7 | 1.3 |  | 1.6 |  |  | 0.4 |  | 1.5 |  | 4.4 | 5.6 | 5.5 | 11.2 | 11.6 |
| 2008 | 0.21 | 3.9 | 1.2 | 0.6 | 0.3 |  | 1.0 |  |  | 0.6 |  | 0.7 |  | 2.4 | 2.7 | 2.6 | 8.8 | 6.1 |
| 2009 | 0.26 | 12.4 | 3.5 |  | 1.0 |  | 3.3 |  |  | 0.8 |  | 1.9 |  | 2.5 | 6.8 | 4.9 | 9.5 | 9.6 |
| 2010 | 0.95 | 7.9 |  |  |  |  | 1.5 |  |  | 0.7 |  | 2.5 |  | 2.7 | 5.1 | 5.6 | 11.0 | 7.1 |
| 2011 | 0.25 | 0.3 |  |  |  |  |  |  |  | 0.4 |  | 0.6 |  | 3.9 | 4.6 | 3.0 | 9.7 | 5.7 |

Table 4.3.5.2. Return rates (\%), by year of smolt migration, of wild Atlantic salmon to 2 SW salmon to North American rivers, 1991 to 2010 . The year 1991 was selected for illustration as it is the first year of the commercial fishery moratorium for the island of Newfoundland.

| Smolt year | USA <br> Narraguagus | Scotia-Fundy |  |  | Gulf |  |  |  | Quebec |  |  |  | NfLD <br> Highlands |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Nashwaak | LaHave | St.Mary's | Margaree | NWMiramichi | SW <br> Miramcihi | Miramichi | à la <br> barbe | Saint Jean | Bec scie | de la <br> Trinite |  |
| 1991 |  |  |  |  |  |  |  |  | 0.6 | 0.9 | 0.4 | 0.6 |  |
| 1992 |  |  |  |  |  |  |  |  | 0.5 | 0.7 | 0.4 | 0.5 |  |
| 1993 |  |  |  |  |  |  |  |  | 0.4 | 0.8 | 0.9 | 0.7 | 1.2 |
| 1994 |  |  |  |  |  |  |  |  |  | 0.9 | 1.5 | 0.7 | 1.4 |
| 1995 |  |  |  |  |  |  |  |  |  | 0.9 | 0.4 | 0.5 | 1.3 |
| 1996 |  |  | 0.3 |  |  |  |  |  |  | 0.4 |  | 0.5 | 0.9 |
| 1997 | 0.84 |  | 0.5 |  |  |  |  |  |  |  |  | 1.1 | 1.2 |
| 1998 | 0.29 | 0.7 | 0.4 |  |  |  |  |  |  | 0.4 |  | 0.7 | 1.1 |
| 1999 | 0.50 | 0.8 | 1.0 | - |  | 1.2 |  |  |  | 0.7 |  | 0.2 | 0.7 |
| 2000 | 0.15 | 0.3 | 0.2 |  |  | 0.5 |  |  |  | 1.2 |  | 0.1 | 0.7 |
| 2001 | 0.83 | 0.9 | 0.6 |  |  | 0.6 | 3.3 | 2.3 |  | 0.9 |  | 0.3 |  |
| 2002 | 0.60 | 1.3 | 0.6 |  | 6.2 | 0.7 | 1.4 | 1.3 |  | 0.9 |  | 0.5 |  |
| 2003 | 1.00 | 1.6 | 0.2 |  | 3.9 | 0.9 | 2.0 | 1.6 |  | 1.4 |  | 0.2 |  |
| 2004 | 0.94 | 1.3 |  |  | 3.0 | 0.5 | 0.8 | 0.7 |  | 1.1 |  | 0.7 |  |
| 2005 | 0.71 | 1.5 | 0.7 | 0.2 | 2.3 | 1.1 |  |  |  | 0.6 |  | 0.5 |  |
| 2006 | 0.74 | 0.6 | 0.4 | 0.1 | 3.0 | 0.2 | 0.5 | 0.4 |  | 0.5 |  |  |  |
| 2007 | 1.99 | 1.3 | 0.2 | 0.0 | 2.1 |  | 0.8 |  |  | 0.5 |  | 0.3 |  |
| 2008 | 0.63 | 2.1 | 0.4 |  | 2.4 |  | 0.7 |  |  | 1.8 |  | 0.5 |  |
| 2009 | 1.71 | 3.3 |  |  | 5.7 |  | 2.2 |  |  | 1.9 |  | 0.8 |  |
| 2010 | 0.20 | 0.0 |  |  |  |  |  |  |  | 1.0 |  | 0.6 |  |
| 2011 |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 4.3.5.3. Return rates (\%), by year of smolt migration, of hatchery Atlantic salmon to 1SW salmon to North American rivers, 1991 to 2011 . The year 1991 was selected for illustration as it is the first year of the commercial fishery moratorium for the island of Newfoundland.

|  | USA |  |  | SF |  |  |  | Gulf |  |  |  | Quebec |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Smolt year | Connecticut | Penobscot | Merrimack | Saint <br> John | LaHave | East <br> Sheet | Liscomb | Morell | Mill | West | Valley-field | auxRochers |
| 1991 | 0.003 | 0.14 | 0.01 | 0.69 | 4.51 | 0.15 | 0.50 | 3.16 |  |  | 0.48 | 0.43 |
| 1992 |  | 0.04 | 0.00 | 0.41 | 1.26 | 0.21 | 0.42 | 1.43 | 0.44 | 2.16 | 0.70 | 0.07 |
| 1993 | 0.003 | 0.05 | 0.00 | 0.39 | 0.62 | 0.32 | 0.56 | 0.14 | 0.37 |  | 0.02 | 0.10 |
| 1994 | 0.003 | 0.03 | 0.00 | 0.66 | 1.44 | 0.36 | 0.35 | 5.20 | 0.11 |  | 0.08 | 0.02 |
| 1995 |  | 0.09 | 0.02 | 1.14 | 2.26 | 0.37 | 0.64 |  |  |  |  | 0.07 |
| 1996 |  | 0.04 | 0.02 | 0.56 | 0.47 | 0.07 | 0.17 |  |  |  |  | 0.31 |
| 1997 |  | 0.04 | 0.02 | 0.75 | 0.87 | 0.03 | 0.15 |  |  |  |  | 0.46 |
| 1998 |  | 0.04 | 0.09 | 0.47 | 0.34 | 0.05 | 0.10 |  |  |  |  | 1.04 |
| 1999 |  | 0.03 | 0.05 | 0.46 | 0.79 | 0.23 |  |  |  |  |  | 0.32 |
| 2000 | 0.003 | 0.03 | 0.01 | 0.27 | 0.43 | 0.03 |  |  |  |  |  | 1.15 |
| 2001 |  | 0.07 | 0.06 | 0.45 | 0.87 |  |  |  |  |  |  | 0.02 |
| 2002 |  | 0.04 | 0.02 | 0.34 | 0.63 |  |  |  |  |  |  | 0.07 |
| 2003 |  | 0.05 | 0.03 | 0.32 | 0.72 |  |  |  |  |  |  |  |
| 2004 |  | 0.05 | 0.02 | 0.39 | 0.53 |  |  |  |  |  |  |  |
| 2005 | 0.015 | 0.06 | 0.02 | 0.56 |  |  |  |  |  |  |  |  |
| 2006 | 0.000 | 0.04 | 0.02 | 0.24 |  |  |  |  |  |  |  |  |
| 2007 | 0.010 | 0.13 | 0.01 | 0.83 |  |  |  |  |  |  |  |  |
| 2008 | 0.000 | 0.03 | 0.00 | 0.13 |  |  |  |  |  |  |  |  |
| 2009 |  | 0.07 | 0.03 | 1.44 |  |  |  |  |  |  |  |  |
| 2010 | 0.005 | 0.12 | 0.18 | 0.12 |  |  |  |  |  |  |  |  |
| 2011 | 0.000 | 0.00 | 0.00 | 0.02 |  |  |  |  |  |  |  |  |

Table 4.3.5.4. Return rates (\%), by year of smolt migration, of hatchery Atlantic salmon to 2SW salmon to North American rivers, 1991 to 2010. The year 1991 was selected for illustration as it is the first year of the commercial fishery moratorium for the island of Newfoundland.

| Smolt year | USA |  |  | SF |  |  |  | Gulf |  |  |  | Quebec |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Connecticut | Penobscot | Merrimack | Saint John | LaHave | East Sheet | Liscomb | Morell | Mill | West | Valley-field | auxRochers |
| 1991 | 0.039 | 0.19 | 0.02 | 0.15 | 0.48 | 0.00 | 0.05 | 0.04 |  |  | 0.00 | 0.13 |
| 1992 | 0.084 | 0.08 | 0.00 | 0.22 | 0.24 | 0.01 | 0.03 | 0.07 | 0.00 | 0.05 | 0.06 | 0.06 |
| 1993 | 0.041 | 0.19 | 0.03 | 0.19 | 0.21 | 0.02 | 0.03 | 0.31 | 0.91 |  | 0.01 | 0.19 |
| 1994 | 0.038 | 0.21 | 0.05 | 0.27 | 0.23 | 0.06 | 0.02 |  |  |  |  | 0.05 |
| 1995 |  | 0.16 | 0.06 | 0.19 | 0.23 | 0.00 | 0.03 |  |  |  |  | 0.04 |
| 1996 |  | 0.14 | 0.09 | 0.08 | 0.13 | 0.01 |  |  |  |  |  | 0.07 |
| 1997 |  | 0.10 | 0.11 | 0.20 | 0.17 | 0.01 |  |  |  |  |  | 0.08 |
| 1998 |  | 0.05 | 0.06 | 0.06 | 0.11 | 0.00 |  |  |  |  |  | 0.09 |
| 1999 |  | 0.08 | 0.13 | 0.16 | 0.21 | 0.00 |  |  |  |  |  | 0.02 |
| 2000 | 0.006 | 0.06 | 0.03 | 0.05 | 0.07 |  |  |  |  |  |  | 0.01 |
| 2001 |  | 0.16 | 0.26 | 0.15 | 0.13 |  |  |  |  |  |  | 0.02 |
| 2002 |  | 0.17 | 0.18 | 0.11 | 0.17 |  |  |  |  |  |  |  |
| 2003 | 0.004 | 0.12 | 0.05 | 0.06 | 0.09 |  |  |  |  |  |  |  |
| 2004 | 0.034 | 0.12 | 0.13 | 0.09 | 0.11 |  |  |  |  |  |  |  |
| 2005 |  | 0.10 | 0.10 | 0.12 |  |  |  |  |  |  |  |  |
| 2006 |  | 0.23 | 0.15 | 0.06 |  |  |  |  |  |  |  |  |
| 2007 |  | 0.30 | 0.08 | 0.17 |  |  |  |  |  |  |  |  |
| 2008 | 0.010 | 0.15 | 0.05 | 0.16 |  |  |  |  |  |  |  |  |
| 2009 | 0.035 | 0.39 | 0.17 | 0.13 |  |  |  |  |  |  |  |  |
| 2010 | 0.002 | 0.09 | 0.11 | 0.07 |  |  |  |  |  |  |  |  |
| 2011 |  |  |  |  |  |  |  |  |  |  |  |  |



Figure 4.1.2.1. Map of Salmon Fishing Areas (SFAs) and Quebec Management Zones (Qs) in Canada.


Figure 4.1.2.2. Summary of recreational fisheries management measures in Canada in 2012.


Figure 4.1.3.1. Harvest ( $\mathbf{t}$ ) of small salmon, large salmon and both sizes combined for Canada, 1960 to 2012 (top panel) and 2003 to 2012 (bottom panel) by all users.



Figure 4.1.3.2. Harvest (number) of small salmon, large salmon and both sizes combined in the recreational fisheries of Canada, 1974 to 2012 (top panel) and 2003 to 2012 (bottom panel).


Figure 4.1.5.1. Exploitation rates in North America on the North American stock complex of small and large salmon, 1971 to 2012.


Figure 4.3.1.1 Time-series of wild smolt production from ten monitored rivers in eastern Canada and one river in eastern USA, 1970 to 2012. Smolt production is expressed as a proportion of the conservation egg requirements for the river.


Figure 4.3.2.1. Comparison of estimated medians of small returns (squares) to and small spawners (circles) in six geographic areas of North America. Returns and spawners for Scotia-Fundy do not include those from SFA 22 and a portion of SFA 23. Note the difference in scale for USA.








Figure 4.3.2.2. Comparison of estimated medians of large returns (squares) to and large spawners (circles) in six geographic areas of North America. Returns and spawners for Scotia-Fundy do not include those from SFA 22 and a portion of SFA 23. For USA estimated spawners exceed the estimated returns due to adult stocking restoration efforts. Also note the difference in scale for USA.


Figure 4.3.2.3. Comparison of the 2 SW conservation limits to the estimated medians of 2 SW returns (squares) and 2SW spawners (circles) in six geographic areas of North America. Returns and spawners for Scotia-Fundy do not include those from SFA 22 and a portion of SFA 23. For USA estimated spawners exceed the estimated returns due to adult stocking restoration efforts. Also note the difference in scale for USA.


Figure 4.3.4.1. Proportion of the conservation requirement attained in the 74 assessed rivers of the North American Commission area in 2012.


Figure 4.3.5.1. Standardized mean (one standard error bars) annual return rates of wild and hatchery origin smolts to 1 SW and 2 SW salmon to the geographic areas of North America. The standardized values are annual means derived from a general linear model analysis of rivers in a region. Survival rates were $\log$ transformed prior to analysis. Note y-scale differences among panels. Error bars are not included for estimates based on a single population.


Figure 4.3.5.2. The percent change in the five-year mean return rates for 1SW and 2SW salmon smolts returning to rivers of eastern North America in 2008 to 2012 compared to the previous period ( 2003 to 2007). Grey circles are for 1 SW and dark squares are for $2 S W$ dataseries. Populations with at least three datapoints in each of the two time periods are included in the analysis.


Figure 4.3.6.2.1. Estimates of PFA for 1SW maturing salmon, 1SW non-maturing salmon and the total cohort of 1SW salmon based on the Monte Carlo simulations of the run-reconstruction model for NAC. Median and $95 \%$ CI interval ranges derived from Monte Carlo simulations are shown.

## 5 Atlantic salmon in the West Greenland Commission

The previous advice provided by ICES (2012c) indicated that there were no catch options for the West Greenland fishery for the years 2012-2014. The NASCO Framework of Indicators for the West Greenland fishery did not indicate the need for a revised analysis of catch options and therefore no new management advice for 2013 is provided. This year's assessment of the contributing stock complexes confirms that advice.

### 5.1 NASCO has requested ICES to describe the events of the 2012 fishery and status of the stocks

### 5.1.1 Catch and effort in 2012

An extant exploitation rate for NAC and NEAC non-maturing 1SW fish at West Greenland can be calculated by dividing the recorded harvest of 1SW salmon at West Greenland by the PFA estimate for the corresponding year for each complex. Exploitation rates are available for the 1971 to 2011 PFA years (Figure 5.1.1.1). The most recent estimate of exploitation available is for the 2011 fishery as the 2012 exploitation rate estimates are dependent on the 2013 returns of 2SW to NAC or MSW to Southern NEAC. NAC PFA estimates are provided for August of the PFA year and NEAC PFA estimates are provided for January of the PFA year, the latter adjusted by eight months (January to August) of natural mortality at 0.03 per month. The 2011 NAC exploitation rate was $7.5 \%$ and is an increase from the previous year's estimate $(5.9 \%)$, is approximate to the previous five-year mean ( $7.3 \%$ ), and remains among the lowest in the time-series. NAC exploitation rate peaked in 1971 at $38.6 \%$. The 2011 NEAC exploitation rate was $0.1 \%$ and is a decrease from the previous year's estimate $(0.5 \%)$, is below the previous five-year mean ( $0.5 \%$ ), and also remains among the lowest in the time-series. NEAC exploitation rate peaked in 1975 at $28.8 \%$.

The Atlantic salmon fishery is currently regulated according to the Government of Greenland Executive Order No 12 of August 1, 2012, which replaces the previous Executive Order no. 21 of August 10, 2002. The only significant change from the previous regulations is that fishermen are no longer required to submit daily catch reports, rather they can record their daily catches in a journal and the journal can be submitted at the end of the season. As before, only hook, fixed gillnets and driftnets are allowed to target salmon directly and the minimum mesh size has been 140 mm (stretched mesh) since 1985. Fishing seasons have varied from year to year, but in general the season has started in August and continued until the quota has been met or until a specified date later in the season. As in recent years, the 2012 season was August 1 to October 31.
Catch data were collated from fisher reports. The reports were screened for errors and missing values. Catches were assigned to NAFO/ICES area based on the reporting community. Reports which contained only the total number of salmon caught or the total catch weight without the number of salmon, were corrected using an average of 3.25 kg gutted weight per salmon. Since 2005 it has been mandatory to report gutted weights, and these have been converted to whole weight using a conversion multiplier of 1.11.

Catches of Atlantic salmon decreased until the closure of the export commercial fishery in 1998, but the subsistence fishery has been increasing in recent years (Table 5.1.1.1; Figure 5.1.1.2). In 2012, catches were distributed among the six NAFO Divi-
sions on the west coast of Greenland and in ICES Division XIV (East Greenland) (Table 5.1.1.2; Figure 5.1.1.3). A total catch of 33.1 t of salmon was reported for the 2012 fishery compared to 27.5 t of salmon in the 2011 fishery, an increase of $20 \%$ from 2011. A harvest of 0.5 t was reported from East Greenland in 2012, accounting for approximately $1.4 \%$ of the total reported catch at Greenland. Harvest reported for east Greenland is not included in assessments of the contributing stock complexes.
With the closure of the commercial fishery since 1998, with the exception of 2001, the export of Atlantic salmon has been banned. From 2002 to 2011, licensed fishermen have been allowed to sell salmon to hotels, institutions and local markets only. In 2012 licensed fishermen were also allowed to land to factories, although the export ban persisted and the landed salmon could only be sold within Greenland.

Reported landings to factories in 2012 occurred in four communities (three communities in NAFO Division 1C and one in NAFO Division 1D) and amounted to 13.7 t . If landings to factories continue in future years, there may be a possibility to place samplers in the communities with factories receiving fish, thereby increasing access to landed fish. Increasing the proportion of sampled fish will improve the characterization of the biological characteristics of the harvest. The Working Group recommends that the Government of Greenland facilitate the coordination of sampling within factories receiving Atlantic salmon, if landings to factories are allowed in 2013.

ICES (2012a) previously identified issues with underreporting of harvest data and it is unclear if the initiation of factory landings and reporting will improve or reduce the accuracy of these data. Landings reported through more centralized sources may help increase the accuracy of the reported landings or over-reporting of landings may occur if fisherman and the factory both make reports.

It is unclear what effect the new landings option had on the effort and total harvest at Greenland. In 2012, a total of 40 fishermen provided 216 reports for a total reported harvest of 19.5 t in the four communities where landings to factories were reported (3.2, 2.7 and 13.7 t for commercial, private and factory landings respectively). In 2011, when there were no landings to factories in these same communities, there were 35 fishermen who provided 151 reports for a total reported harvest of 13.3 t . This represents an increase in five fishermen, 65 reports and 5.8 t of landings from 2011 vs .2012. Additionally, two of the communities that reported factory landings in 2012 had a total of 20 salmon reported in 2011. However, 15 t of landings were reported for Division 1 C , of which 12 t were landings to factories. This is the highest reported landings for Division 1C since 1997. The increase in reported landings could be due to increased effort, catch per unit of effort or reporting rates.

There is currently no quantitative approach for estimating the unreported catch but the 2012 value is likely to have been at the same level proposed in recent years ( 10 t ).

Of the total catch, 14.1 t was reported as being for private consumption, 5.5 t as commercial and 13.7 t as factory landings. However, 9.9 t of the private consumption catch was reported by licensed fishers.

In recent years there seems to have been almost no transport of salmon from the settlements to the cities and instead, the fisheries have been conducted in close proximity to the cities. Additionally, commercial catches are highest in the cities and lowest around settlements where the private fishery is more prevalent. This dynamic may change if landings to factories continue and salmon are shipped from processing plants to various communities for sale.

The seasonal distribution of catches has previously been reported to the Working Group (ICES 2002). However since 2002, this has not been possible. Although fishers are required to record daily catches comparisons of summed reported catch and number of returned catch reports reveals that a large number of fishers report their total catch in only one report for the entire season.

The Greenland Authorities received 553 reports of salmon catches from 122 fishers in 2012 compared to 394 reports from 117 fishers in 2011 (Table 5.1.1.3). The number of fishers was similar although the number of reports increased significantly. The total number of fishers reporting catches from all areas has increased from a low of 41 in 2002 to its current level. These levels remain well below the 400 to 600 people reporting landings in the commercial export fishery from 1987 to 1991.

The variations in the numbers of people reporting catches as well as the catches in each of the NAFO Divisions suggest that there are inconsistencies in the catch data and highlights the need for better data. Continuation and improvement of the voluntary logbook reporting system initiated by the Greenlandic Authorities in 2011 will help to improve the quality of the reported catch statistics.

The logbook instructions requested that fishers provide more detailed information related to their salmon fishing activities than required under their licence conditions. Logbooks were only provided to fishers holding commercial licences, and not to private fishers. The data requested were:

- Date;
- Fishing place;
- Number of salmon;
- Weight in kg (gutted);
- Number of nets;
- Number of fishing hours;
- Catch sold/Community catch sold in;
- Notes.

It was noted that factory landing reports contain similar information to that requested in the logbooks. These data will allow for a more accurate characterization of the nature and extent of the fishery than is currently available. Logbook and factory data may provide catch and effort statistics (cpue) that will allow a more detailed assessment based on time and location of fishing activities. More detailed information on the nature and extent of the fishery will allow for better management of this resource. Catch per unit of effort (cpue) statistics represent indirect measures of the abundance and trends. Increasing cpue values may be indicative of increasing abundance, decreasing cpue values may be indicative of decreasing abundance, and constant cpue values may be indicative of stable abundance.

The Working Group recommends that the reporting system continues and that logbooks be provided to all fishers. Efforts should continue to encourage compliance with the logbook voluntary system. Detailed statistics related to catch and effort should be made available to the Working Group for analysis.

### 5.1.2 Biological characteristics of the catches

The international sampling programme for the fishery at West Greenland agreed by the parties at NASCO continued in 2012. The sampling was undertaken by participants from Canada, Ireland, UK(Scotland), UK(England\&Wales), and USA. Sampling
began in August and continued through October. Additionally, staff from the Greenland Institute of Natural Resources assisted with coordination of the programme.
Samplers were stationed in three different communities (Figure 5.1.1.3) representing three different NAFO Divisions: Sisimiut (1B), Maniitsoq (1C), and Qaqortoq (1F). As in previous years no sampling occurred in the fishery in East Greenland. In this Baseline Sampling Programme, tissue and biological samples were collected.
In total 1378 individual salmon were sampled representing $\sim 14 \%$ by weight of the reported landings. Of these, 1372 fork lengths were measured (Table 5.1.2.1). Scale samples were taken from 1371 salmon for age determination and 1373 tissue samples were collected for DNA analysis and continent of origin assignment.

A total of 17 adipose finclipped fish were recovered, but none of these carried tags. However, a total of six tags were recovered during the fishing season: no tags were recovered by the sampling programme and all six tags were returned directly to the Nature Institute. Five tags came from Canadian origin fish: four from incoming 'bright' adult salmon tagged and released in the Mirimachi River (Gulf Region) in 2011 and one from an outmigrating kelt tagged and released on the Campbellton River (Newfoundland) in 2012. One tag came from Sweden: a smolt tagged and released in 2011.

In all years since 2002, except for 2006 and in 2011, non-reporting of harvest was evident based on a comparison of reported landings to the sample data. In at least one of the NAFO Divisions where international samplers were present, the sampling team observed more fish than were reported as being landed. When there is this type of weight discrepancy, the reported landings are adjusted according to the total weight of the fish identified as being landed during the sampling effort and these adjusted landings are carried forward for all future assessments. The time-series of reported landings and subsequent adjusted landings for 2002-2012 are presented in Table 5.1.2.2. The 2012 adjusted landings represented a 2 t increased over the reported landings.

As reported previously (ICES, 2012a), access to fish in support of the Baseline Sampling Programme in Nuuk has been compromised. No solution to this issue was reached prior to the 2012 sampling season and consequently no sampling was conducted within the capital city. Unless assurances can be provided that access to fish will be allowed, sampling in Nuuk may not occur for the foreseeable future.
The small catch levels and the broad geographic and temporal coverage of the internal use only fishing caused severe practical problems for the sampling teams. Despite these constraints, the sampling programme successfully sampled the Greenland catch, both temporally and spatially. The need to obtain samples from fish landed in Nuuk, should be reiterated. Nuuk accounted for $13 \%$ of the adjusted landings in 2012, $29 \%$ in 2011 and 19\% for the period 2003-2012 (range 12-29\%). Not being able to sample fish landed in Nuuk may compromise the sampling programme's ability to collect the samples needed to accurately describe the biological characteristics of the salmon harvest at West Greenland. The Working Group recommends that arrangements be made to enable sampling in Nuuk as a significant amount of salmon is reported as being landed in this community on an annual basis.

The mean length and whole weight of North American 1SW salmon was 65.5 cm and 3.34 kg weight and the means for European 1 SW salmon were 64.9 cm and 3.38 kg (Table 5.1.2.3). The North American and European 1SW whole weight estimates increased slightly from 2011 values and are greater than the ten year mean. The North

American 1SW fork length remained approximately the same as the 2011 estimate and was greater than the ten year mean. The European 1SW fork length remained approximately the same as the 2011 value and the ten year mean.

Over the period of sampling (1969 to 2012) the mean weight of 1SW non-maturing salmon at West Greenland declined from high values in the 1970s to the lowest mean weights of the time-series in 1990 to 1995, before increasing subsequently to 2010. Mean weight have since remained close to the 2010 level. However, these mean weight trends are unadjusted for the period of sampling and it is known that salmon grow quickly during the period of sampling in the fishery from August to October.
The Working Group previously examined the changing weights and condition factors of 1SW non-maturing salmon at West Greenland (ICES, 2011b). The analysis of condition of salmon over the period 2002 to 2010 (time period of data available) contrasts with the interpretation of salmon size at West Greenland based entirely on weights or lengths unadjusted for the period of sampling or for the length of the fish (ICES, 2011b). With few exceptions, there was no apparent change in condition of 1SW nonmaturing salmon at West Greenland. The trend in increasing weights from the samples can be attributed to both increasing length and variations in sampling period.

The Working Group recommends that the longer time-series of sampling data from West Greenland should be analysed to assess the extent of the variations in condition over the time period corresponding to the large variations in productivity as identified by the NAC and NEAC assessment and forecast models. Progress has been made compiling the West Greenland sampling database and should be available for analysis prior to the 2014 Working Group meeting.

North American salmon up to river age six years were sampled from the fishery at West Greenland (Table 5.1.2.4), comprised predominantly of two year old (29.8\%), three year old (39.4\%) and four year old (23.3\%) smolts. The river ages of European salmon ranged from one to four years (Table 5.1.2.5). Of these, $9.3 \%$ were river age one, $63.0 \%$ were river age two, and $24.0 \%$ were river age three.

As expected, the 1SW age group dominated the 2012 sample collection for both the North American and European origin fish ( $93.2 \%$ and $98.0 \%$ respectively, Table 5.1.2.6).

As part of the sampling programme sex was determined by gonadal examination of only 16 salmon. They were comprised of $25.0 \%$ males and $75.0 \%$ females.

### 5.1.3 Continent of origin of catches at West Greenland

A total of 1373 samples were collected from salmon from three communities representing three NAFO Divisions: Sisimiut (1B, $n=464$ ), Maniitsoq (1C, $n=585$ ), and Qaqortoq ( $1 \mathrm{~F}, \mathrm{n}=324$ ). DNA isolation and the subsequent microsatellite analysis were performed (King et al., 2001). As in previous years, a database of approximately 5000 Atlantic salmon genotypes of known origin was used as a baseline to assign these individuals to continent of origin. In total, $81.6 \%$ of the salmon sampled were of North American origin and $18.4 \%$ were determined to be of European origin. The NAFO Division-specific continent of origin assignments are presented in Table 5.1.3.1.

These data show the high proportion of North American origin individuals contributing to the fishery over the recent past (Table 5.1.3.2; Figure 5.1.3.1). The variability in the recent continental representation among divisions (Table 5.1.3.1) underscores
the need to sample multiple NAFO Divisions to achieve the most accurate estimate of the contribution of fish from each continent to the mixed-stock fishery.

The estimated weighted proportions of North American and European salmon since 1982 and the weighted numbers of North American and European Atlantic salmon caught at West Greenland (excluding the reported harvest from ICES Area XIV) are provided in Table 5.1.3.2 and Figure 5.1.3.2. Approximately 7800 ( $\sim 27.2$ t) North American origin fish and approximately $2100(\sim 7.3 \mathrm{t})$ European origin fish were harvested in 2012. These remain among the lowest in the time-series, but the second highest in the past ten years (2003-2012).

The Working Group recommends a continuation and expansion of the broad geographic sampling programme (multiple NAFO divisions) to more accurately estimate continent of origin in the mixed-stock fishery.

### 5.2 NASCO has requested ICES to describe the status of the stocks

Six out of the seven stock complexes exploited at West Greenland are below conservation limits. In European and North American areas, the overall abundance of stocks contributing to the West Greenland fishery has recently increased, however the abundance of salmon within the West Greenland area remains low relative to historical levels. A more detailed overview of status of stocks in the NEAC and NAC areas is presented in the relevant commission sections (Sections 3 and 4).

### 5.2.1 North American stock complex

North American 2SW spawner estimates were below their CLs in all six regions (Figure 4.3.2.3) in 2012. Within each of the geographic areas there are varying numbers of individual river stocks which are failing to meet CLs, particularly in the southern areas of Scotia-Fundy and the USA. The estimated exploitation rate of North American origin salmon in North American fisheries has declined (Figure 4.1.5.1) from approximately $68 \%$ in 1973 to $14.3 \%$ in 2012 for 1SW salmon and over $84 \%$ in 1981 to $11.9 \%$ in 2012 for 2 SW salmon. The 2012 exploitation rates on 1SW and 2SW salmon both remained close to the 2011 estimates ( $14.9 \%$ and $10.6 \%$ respectively) and among the lowest in the time-series.

### 5.2.2 Southern European stock complex

The status of stocks in the four Northeast Atlantic stock complexes is assessed with respect to abundance relative to spawning escapement reserve and prior to the commencement of distant water fisheries. All four stock complexes (Northern NEAC 1SW and MSW and Southern NEAC 1SW and MSW) were considered to be at full reproductive capacity prior to the commencement of distant-water fisheries, in the latest available PFA year (2011) (Figure 3.3.5.1). However, at a country level, stocks from several jurisdictions were below CLs (Figures 3.3.4.1.a-j). Stocks from countries in Northern NEAC area were generally above their CLs while stocks from countries in Southern NEAC were generally below their CLs. Further, within all countries there were individual river stocks that are not meeting CLs (Table 3.3.5.1). Exploitation rates on these four stock complexes (Northern NEAC 1SW and MSW and Southern NEAC MSW) are shown in Figures 3.1.9.1 and 3.1.9.2. Exploitation rates on 1SW salmon in the Northern and Southern NEAC areas were $40 \%$ and $12 \%$ in 2012; both representing declines from the previous five year averages ( $41 \%$ and $15 \%$ respectively). Exploitation rates on MSW salmon in the Northern and Southern NEAC areas were $44 \%$ and $11 \%$ in 2012; both representing declines from the previous five year
averages ( $49 \%$ and $13 \%$ respectively). These current estimates for both stock complexes are at or among the lowest in the time-series.

Table 5.1.1.1. Nominal catches of salmon at West Greenland since 1971 (metric tons round fresh weight).

| Year | Total | Quota | Comments |
| :---: | :---: | :---: | :---: |
| 1971 | 2689 | - |  |
| 1972 | 2113 | 1100 |  |
| 1973 | 2341 | 1100 |  |
| 1974 | 1917 | 1191 |  |
| 1975 | 2030 | 1191 |  |
| 1976 | 1175 | 1191 |  |
| 1977 | 1420 | 1191 |  |
| 1978 | 984 | 1191 |  |
| 1979 | 1395 | 1191 |  |
| 1980 | 1194 | 1191 |  |
| 1981 | 1264 | 1265 | Quota set to a specific opening date for the fishery |
| 1982 | 1077 | 1253 | Quota set to a specific opening date for the fishery |
| 1983 | 310 | 1191 |  |
| 1984 | 297 | 870 |  |
| 1985 | 864 | 852 |  |
| 1986 | 960 | 909 |  |
| 1987 | 966 | 935 |  |
| 1988 | 893 | 840 | Quota for 1988-90 was 2520 t with an opening date of |
| 1989 | 337 | 900 | were not to exceed an annual |
| 1990 | 274 | 924 | in 1989 and 924 t in 1990 for later opening dates. |
| 1991 | 472 | 840 |  |
| 1992 | 237 | 258 | Quota set by Greenland authorities |
| 1993 |  | 89 | The fishery was suspended. NASCO adopt a new quota allocation model. |
| 1994 |  | 137 | The fishery was suspended and the quotas were bought out. |
| 1995 | 83 | 77 | Quota advised by NASCO |
| 1996 | 92 | 174 | Quota set by Greenland authorities |
| 1997 | 58 | 57 | Private (non-commercial) catches to be reported from now |
| 1998 | 11 | 20 | Fishery restricted to catches used for internal |
| 1999 | 19 | 20 | consumption in Greenland |
| 2000 | 21 | 20 |  |
| 2001 | 43 | 114 | Final quota calculated according to the ad hoc management system |
| 2002 | 9 | 55 | Quota bought out, quota represented the maximum allowable catch (no factory landing allowed), and higher catch figures based on sampling programme information are used for the assessments |
| 2003 | 9 |  | Quota set to nil (no factory landing allowed), fishery restricted to catches used for internal consumption in Greenland, and higher catch figures based on sampling programme information are used for the assessments |
| 2004 | 15 |  | same as previous year |


| Year | Total | Quota |
| :---: | :---: | :--- |
| 2005 | 15 | Comments |
| 2006 | 22 | same as previous year <br> Quota set to nil (no factory landing allowed) and fishery <br> restricted to catches used for internal consumption in <br> Greenland |
| 2007 | 25 | Quota set to nil (no factory landing allowed), fishery <br> restricted to catches used for internal consumption in <br> Greenland, and higher catch figures based on sampling <br> programme information are used for the assessments |
| 2008 | 26 | same as previous year |
| 2009 | 26 | same as previous year |
| 2010 | 40 | same as previous year |
| 2011 | 28 | same as previous year |
| 2012 | 33 | Quota set to nil (factory landing allowed), fishery <br> restricted to catches used for internal consumption in <br> Greenland, and higher catch figures based on sampling <br> programme information are used for the assessments |

Table 5.1.1.2. Distribution of nominal catches (metric tons) by Greenland vessels since 1977. NAFO Division is represented by $1 \mathrm{~A}-1 \mathrm{~F}$.

| Year | 1 A | 1 B | 1 C | 1D | 1E | 1 F | Unk. | West Greenland | East <br> Greenland | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 201 | 393 | 336 | 207 | 237 | 46 | - | 1420 | 6 | 1426 |
| 1978 | 81 | 349 | 245 | 186 | 113 | 10 | - | 984 | 8 | 992 |
| 1979 | 120 | 343 | 524 | 213 | 164 | 31 | - | 1395 | + | 1395 |
| 1980 | 52 | 275 | 404 | 231 | 158 | 74 | - | 1194 | + | 1194 |
| 1981 | 105 | 403 | 348 | 203 | 153 | 32 | 20 | 1264 | + | 1264 |
| 1982 | 111 | 330 | 239 | 136 | 167 | 76 | 18 | 1077 | + | 1077 |
| 1983 | 14 | 77 | 93 | 41 | 55 | 30 | - | 310 | + | 310 |
| 1984 | 33 | 116 | 64 | 4 | 43 | 32 | 5 | 297 | + | 297 |
| 1985 | 85 | 124 | 198 | 207 | 147 | 103 | - | 864 | 7 | 871 |
| 1986 | 46 | 73 | 128 | 203 | 233 | 277 | - | 960 | 19 | 979 |
| 1987 | 48 | 114 | 229 | 205 | 261 | 109 | - | 966 | + | 966 |
| 1988 | 24 | 100 | 213 | 191 | 198 | 167 | - | 893 | 4 | 897 |
| 1989 | 9 | 28 | 81 | 73 | 75 | 71 | - | 337 | - | 337 |
| 1990 | 4 | 20 | 132 | 54 | 16 | 48 | - | 274 | - | 274 |
| 1991 | 12 | 36 | 120 | 38 | 108 | 158 | - | 472 | 4 | 476 |
| 1992 | - | 4 | 23 | 5 | 75 | 130 | - | 237 | 5 | 242 |
| $1993{ }^{1}$ | - | - | - | - | - | - | - | - | - | - |
| $1994{ }^{1}$ | - | - | - | - | - | - | - | - | - | - |
| 1995 | + | 10 | 28 | 17 | 22 | 5 | - | 83 | 2 | 85 |
| 1996 | + | + | 50 | 8 | 23 | 10 | - | 92 | + | 92 |
| 1997 | 1 | 5 | 15 | 4 | 16 | 17 | - | 58 | 1 | 59 |
| 1998 | 1 | 2 | 2 | 4 | 1 | 2 | - | 11 | - | 11 |
| 1999 | + | 2 | 3 | 9 | 2 | 2 | - | 19 | + | 19 |
| 2000 | + | + | 1 | 7 | + | 13 | - | 21 | - | 21 |
| 2001 | + | 1 | 4 | 5 | 3 | 28 | - | 43 | - | 43 |
| 2002 | + | + | 2 | 4 | 1 | 2 | - | 9 | - | 9 |
| 2003 | 1 | + | 2 | 1 | 1 | 5 | - | 9 | - | 9 |
| 2004 | 3 | 1 | 4 | 2 | 3 | 2 | - | 15 | - | 15 |
| 2005 * | 1 | 3 | 2 | 1 | 3 | 5 | - | 15 | - | 15 |
| 2006 * | 6 | 2 | 3 | 4 | 2 | 4 | - | 22 | - | 22 |
| 2007 * | 2 | 5 | 6 | 4 | 5 | 2 | - | 25 | - | 25 |
| 2008* | 4.9 | 2.2 | 10.0 | 1.6 | 2.5 | 5.0 | 0 | 26.2 | 0 | 26.2 |
| 2009 * | 0.2 | 6.2 | 7.1 | 3.0 | 4.3 | 4.8 | 0 | 25.6 | 0.8 | 26.4 |
| 2010* | 17.3 | 4.6 | 2.4 | 2.7 | 6.8 | 4.3 | 0 | 38.1 | 1.7 | 39.6 |
| 2011* | 1.8 | 3.7 | 5.3 | 8.0 | 4.0 | 4.6 | 0 | 27.4 | 0.1 | 27.5 |
| 2012* | 5.4 | 0.8 | 15.0 | 4.6 | 4.0 | 3.0 | 0 | 32.6 | 0.5 | 33.1 |

${ }^{1}$ The fishery was suspended.

+ Small catches $<5$ t.
- No catch.
* Corrected from gutted weight to total weight (factor 1.11).

Table 5.1.1.3. Number of people (licensed and unlicensed) reporting catches of Atlantic salmon in the Greenland fishery and the total number of licences issued by NAFO (1A-1F)/ICES Divisions. Reports received by fish plants prior to 1997 and to the Licence Office from 1998 to present.

| Year | 1 A | 1 B | 1 C | 1D | 1E | 1 F | ICES | Unk. | Licences | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 78 | 67 | 74 |  | 99 | 233 |  | 0 |  | 579 |
| 1988 | 63 | 46 | 43 | 53 | 78 | 227 |  | 0 |  | 516 |
| 1989 | 30 | 41 | 98 | 46 | 46 | 131 |  | 0 |  | 393 |
| 1990 | 32 | 15 | 46 | 52 | 54 | 155 |  | 0 |  | 362 |
| 1991 | 53 | 39 | 100 | 41 | 54 | 123 |  | 0 |  | 410 |
| 1992 | 3 | 9 | 73 | 9 | 36 | 82 |  | 0 |  | 212 |
| 1993 |  |  |  |  |  |  |  |  |  |  |
| 1994 |  |  |  |  |  |  |  |  |  |  |
| 1995 | 0 | 17 | 52 | 21 | 24 | 31 |  | 0 |  | 145 |
| 1996 | 1 | 8 | 74 | 15 | 23 | 42 |  | 0 |  | 163 |
| 1997 | 0 | 16 | 50 | 7 | 2 | 6 |  | 0 |  | 80 |
| 1998 | 16 | 5 | 8 | 7 | 3 | 30 |  | 0 |  | 69 |
| 1999 | 3 | 8 | 24 | 18 | 21 | 29 |  | 0 |  | 102 |
| 2000 | 1 | 1 | 5 | 12 | 2 | 25 |  | 0 |  | 43 |
| 2001 | 2 | 7 | 13 | 15 | 6 | 37 |  | 0 | 452 | 76 |
| 2002 | 1 | 1 | 9 | 13 | 9 | 8 |  | 0 | 479 | 41 |
| 2003 | 11 | 1 | 4 | 4 | 12 | 10 |  | 0 | 150 | 42 |
| 2004 | 20 | 2 | 8 | 4 | 20 | 12 |  | 0 | 155 | 66 |
| 2005 | 11 | 7 | 17 | 5 | 17 | 18 |  | 0 | 185 | 75 |
| 2006 | 43 | 14 | 17 | 20 | 17 | 30 |  | 0 | 159 | 141 |
| 2007 | 29 | 12 | 26 | 10 | 33 | 22 |  | 0 | 260 | 132 |
| 2008 | 44 | 8 | 41 | 10 | 16 | 24 |  | 0 | 260 | 143 |
| 2009 | 19 | 11 | 35 | 15 | 25 | 31 | 9 | 0 | 294 | 145 |
| 2010 | 86 | 17 | 19 | 16 | 30 | 27 | 13 | 0 | 309 | 208 |
| 2011 | 25 | 9 | 20 | 15 | 20 | 23 | 5 | 0 | 234 | 117 |
| 2012 | 35 | 9 | 32 | 8 | 16 | 16 | 6 | 0 | 279 | 122 |

Table 5.1.2.1. Size of biological samples and percentage (by number) of North American and European salmon in research vessel catches at West Greenland (1969 to 1982), from commercial samples (1978 to 1992, 1995 to 1997, and 2001) and from local consumption samples (1998 to 2000, and 2002 to present).

|  |  | Sample Size |  |  | Continent of Origin (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Source |  |  |  |  |  | (95\% |  | (95\% |
|  |  | Length | Scales | Genetics | NA | CI) ${ }^{1}$ | E | CI) ${ }^{1}$ |
| Research | 1969 | 212 | 212 |  | 51 | $(57,44)$ | 49 | $(56,43)$ |
|  | 1970 | 127 | 127 |  | 35 | $(43,26)$ | 65 | $(75,57)$ |
|  | 1971 | 247 | 247 |  | 34 | $(40,28)$ | 66 | $(72,50)$ |
|  | 1972 | 3488 | 3488 |  | 36 | $(37,34)$ | 64 | $(66,63)$ |
|  | 1973 | 102 | 102 |  | 49 | $(59,39)$ | 51 | $(61,41)$ |
|  | 1974 | 834 | 834 |  | 43 | $(46,39)$ | 57 | $(61,54)$ |
|  | 1975 | 528 | 528 |  | 44 | $(48,40)$ | 56 | $(60,52)$ |
|  | 1976 | 420 | 420 |  | 43 | $(48,38)$ | 57 | $(62,52)$ |
|  | $1978{ }^{2}$ | 606 | 606 |  | 38 | $(41,38)$ | 62 | $(66,59)$ |
|  | $1978{ }^{3}$ | 49 | 49 |  | 55 | $(69,41)$ | 45 | $(59,31)$ |
|  | 1979 | 328 | 328 |  | 47 | $(52,41)$ | 53 | $(59,48)$ |
|  | 1980 | 617 | 617 |  | 58 | $(62,54)$ | 42 | $(46,38)$ |
|  | 1982 | 443 | 443 |  | 47 | $(52,43)$ | 53 | $(58,48)$ |
| Commercial | 1978 | 392 | 392 |  | 52 | $(57,47)$ | 48 | $(53,43)$ |
|  | 1979 | 1653 | 1653 |  | 50 | $(52,48)$ | 50 | $(52,48)$ |
|  | 1980 | 978 | 978 |  | 48 | $(51,45)$ | 52 | $(55,49)$ |
|  | 1981 | 4570 | 1930 |  | 59 | $(61,58)$ | 41 | $(42,39)$ |
|  | 1982 | 1949 | 414 |  | 62 | $(64,60)$ | 38 | $(40,36)$ |
|  | 1983 | 4896 | 1815 |  | 40 | $(41,38)$ | 60 | $(62,59)$ |
|  | 1984 | 7282 | 2720 |  | 50 | $(53,47)$ | 50 | $(53,47)$ |
|  | 1985 | 13272 | 2917 |  | 50 | $(53,46)$ | 50 | $(52,34)$ |
|  | 1986 | 20394 | 3509 |  | 57 | $(66,48)$ | 43 | $(52,34)$ |
|  | 1987 | 13425 | 2960 |  | 59 | $(63,54)$ | 41 | $(46,37)$ |
|  | 1988 | 11047 | 2562 |  | 43 | $(49,38)$ | 57 | $(62,51)$ |
|  | 1989 | 9366 | 2227 |  | 56 | $(60,52)$ | 44 | $(48,40)$ |
|  | 1990 | 4897 | 1208 |  | 75 | $(79,70)$ | 25 | $(30,21)$ |
|  | 1991 | 5005 | 1347 |  | 65 | $(69,61)$ | 35 | $(39,31)$ |
|  | 1992 | 6348 | 1648 |  | 54 | $(57,50)$ | 46 | $(50,43)$ |
|  | 1995 | 2045 | 2045 |  | 68 | $(75,65)$ | 32 | $(35,28)$ |
|  | 1996 | 3341 | 1397 |  | 73 | $(76,71)$ | 27 | $(29,24)$ |
|  | 1997 | 794 | 282 |  | 80 | $(84,75)$ | 20 | $(25,16)$ |
| Local |  |  |  |  |  |  |  |  |
| Consumption | 1998 | 540 | 406 |  | 79 | $(84,73)$ | 21 | $(27,16)$ |
|  | 1999 | 532 | 532 |  | 90 | $(97,84)$ | 10 | $(16,3)$ |
|  | 2000 | 491 | 491 |  | 70 |  | 30 |  |
| Commercial | 2001 | 4721 | 2655 |  | 69 | $(71,67)$ | 31 | $(33,29)$ |


|  |  | Sample Size |  |  | Continent of Origin (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Source |  | Length | Scales | Genetics | NA | $\begin{aligned} & (95 \% \\ & \mathrm{CI})^{1} \end{aligned}$ | E | $\begin{aligned} & (95 \% \\ & \text { CI })^{1} \\ & \hline \end{aligned}$ |
| Local |  |  |  |  |  |  |  |  |
| Consumption | 2002 | 501 | 501 | 501 | 68 |  | 32 |  |
|  | 2003 | 1743 | 1743 | 1779 | 68 |  | 32 |  |
|  | 2004 | 1639 | 1639 | 1688 | 73 |  | 27 |  |
|  | 2005 | 767 | 767 | 767 | 76 |  | 24 |  |
|  | 2006 | 1209 | 1209 | 1193 | 72 |  | 28 |  |
|  | 2007 | 1116 | 1110 | 1123 | 82 |  | 18 |  |
|  | 2008 | 1854 | 1866 | 1853 | 86 |  | 14 |  |
|  | 2009 | 1662 | 1683 | 1671 | 91 |  | 9 |  |
|  | 2010 | 1261 | 1265 | 1240 | 80 |  | 20 |  |
|  | 2011 | 967 | 965 | 964 | 92 |  | 8 |  |
|  | 2012 | 1372 | 1371 | 1373 | 82 |  | 18 |  |

${ }^{1}$ CI - confidence interval calculated by method of Pella and Robertson (1979) for 1984-86 and binomial distribution for the others.
${ }^{2}$ During 1978 Fishery
${ }^{3}$ Research samples after 1978 fishery closed.

Table 5.1.2.2. Reported landings (kg) for the West Greenland Atlantic salmon fishery from 2002 by NAFO Division and the division-specific adjusted landings where the sampling teams observed more fish landed than were reported. Adjusted landings were not calculated for 2006 and 2011 as the sampling teams did not observe more fish than were reported.

| Year |  | IA | 1B | IC | 1D | IE | IF | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2002 | Reported | 14 | 78 | 2100 | 3752 | 1417 | 1661 | 9022 |
|  | Adjusted |  |  |  |  |  | 2408 | 9769 |
| 2003 | Reported | 619 | 17 | 1621 | 648 | 1274 | 4516 | 8694 |
|  | Adjusted |  |  | 1782 | 2709 |  | 5912 | 12312 |
| 2004 | Reported | 3476 | 611 | 3516 | 2433 | 2609 | 2068 | 14712 |
|  | Adjusted |  |  |  | 4929 |  |  | 17209 |
| 2005 | Reported | 1294 | 3120 | 2240 | 756 | 2937 | 4956 | 15303 |
|  | Adjusted |  |  |  | 2730 |  |  | 17276 |
| 2006 | Reported | 5427 | 2611 | 3424 | 4731 | 2636 | 4192 | 23021 |
|  | Adjusted |  |  |  |  |  |  |  |
| 2007 | Reported | 2019 | 5089 | 6148 | 4470 | 4828 | 2093 | 24647 |
|  | Adjusted |  |  |  |  |  | 2252 | 24806 |
| 2008 | Reported | 4882 | 2210 | 10024 | 1595 | 2457 | 4979 | 26147 |
|  | Adjusted |  |  |  | 3577 |  | 5478 | 28627 |
| 2009 | Reported | 195 | 6151 | 7090 | 2988 | 4296 | 4777 | 25496 |
|  | Adjusted |  |  |  | 5466 |  |  | 27975 |
| 2010 | Reported | 17263 | 4558 | 2363 | 2747 | 6766 | 4252 | 37949 |
|  | Adjusted |  | 4824 |  | 6566 |  | 5274 | 43056 |
| 2011 | Reported | 1858 | 3662 | 5274 | 7977 | 4021 | 4613 | 27407 |
|  | Adjusted |  |  |  |  |  |  |  |
| 2012 | Reported | 5353 | 784 | 14991 | 4564 | 3993 | 2951 | 32636 |
|  | Adjusted |  | 2001 |  |  |  | 3694 | 34596 |

Table 5.1.2.3. Annual mean whole weights (kg) and fork lengths ( cm ) by sea age and continent of origin of Atlantic salmon caught at West Greenland 1969 to 1992 and 1995 to present (NA = North America and E = Europe).

|  | Whole weight (kg) |  |  |  |  |  | Fork Length (cm) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW |  | 2SW |  | PS |  | All s | ges | Total | 1SW |  | 2SW |  | PS |  |
|  | NA | E | NA | E | NA | E | NA | E |  | NA | E | NA | E | NA | E |
| 1969 | 3.12 | 3.76 | 5.48 | 5.80 | - | 5.13 | 3.25 | 3.86 | 3.58 | 65.0 | 68.7 | 77.0 | 80.3 | - | 75.3 |
| 1970 | 2.85 | 3.46 | 5.65 | 5.50 | 4.85 | 3.80 | 3.06 | 3.53 | 3.28 | 64.7 | 68.6 | 81.5 | 82.0 | 78.0 | 75.0 |
| 1971 | 2.65 | 3.38 | 4.30 | - | - | - | 2.68 | 3.38 | 3.14 | 62.8 | 67.7 | 72.0 | - | - | - |
| 1972 | 2.96 | 3.46 | 5.85 | 6.13 | 2.65 | 4.00 | 3.25 | 3.55 | 3.44 | 64.2 | 67.9 | 80.7 | 82.4 | 61.5 | 69.0 |
| 1973 | 3.28 | 4.54 | 9.47 | 10.00 | - | - | 3.83 | 4.66 | 4.18 | 64.5 | 70.4 | 88.0 | 96.0 | 61.5 | - |
| 1974 | 3.12 | 3.81 | 7.06 | 8.06 | 3.42 | - | 3.22 | 3.86 | 3.58 | 64.1 | 68.1 | 82.8 | 87.4 | 66.0 | - |
| 1975 | 2.58 | 3.42 | 6.12 | 6.23 | 2.60 | 4.80 | 2.65 | 3.48 | 3.12 | 61.7 | 67.5 | 80.6 | 82.2 | 66.0 | 75.0 |
| 1976 | 2.55 | 3.21 | 6.16 | 7.20 | 3.55 | 3.57 | 2.75 | 3.24 | 3.04 | 61.3 | 65.9 | 80.7 | 87.5 | 72.0 | 70.7 |
| 1978 | 2.96 | 3.50 | 7.00 | 7.90 | 2.45 | 6.60 | 3.04 | 3.53 | 3.35 | 63.7 | 67.3 | 83.6 | - | 60.8 | 85.0 |
| 1979 | 2.98 | 3.50 | 7.06 | 7.60 | 3.92 | 6.33 | 3.12 | 3.56 | 3.34 | 63.4 | 66.7 | 81.6 | 85.3 | 61.9 | 82.0 |
| 1980 | 2.98 | 3.33 | 6.82 | 6.73 | 3.55 | 3.90 | 3.07 | 3.38 | 3.22 | 64.0 | 66.3 | 82.9 | 83.0 | 67.0 | 70.9 |
| 1981 | 2.77 | 3.48 | 6.93 | 7.42 | 4.12 | 3.65 | 2.89 | 3.58 | 3.17 | 62.3 | 66.7 | 82.8 | 84.5 | 72.5 | - |
| 1982 | 2.79 | 3.21 | 5.59 | 5.59 | 3.96 | 5.66 | 2.92 | 3.43 | 3.11 | 62.7 | 66.2 | 78.4 | 77.8 | 71.4 | 80.9 |
| 1983 | 2.54 | 3.01 | 5.79 | 5.86 | 3.37 | 3.55 | 3.02 | 3.14 | 3.10 | 61.5 | 65.4 | 81.1 | 81.5 | 68.2 | 70.5 |
| 1984 | 2.64 | 2.84 | 5.84 | 5.77 | 3.62 | 5.78 | 3.20 | 3.03 | 3.11 | 62.3 | 63.9 | 80.7 | 80.0 | 69.8 | 79.5 |
| 1985 | 2.50 | 2.89 | 5.42 | 5.45 | 5.20 | 4.97 | 2.72 | 3.01 | 2.87 | 61.2 | 64.3 | 78.9 | 78.6 | 79.1 | 77.0 |
| 1986 | 2.75 | 3.13 | 6.44 | 6.08 | 3.32 | 4.37 | 2.89 | 3.19 | 3.03 | 62.8 | 65.1 | 80.7 | 79.8 | 66.5 | 73.4 |
| 1987 | 3.00 | 3.20 | 6.36 | 5.96 | 4.69 | 4.70 | 3.10 | 3.26 | 3.16 | 64.2 | 65.6 | 81.2 | 79.6 | 74.8 | 74.8 |
| 1988 | 2.83 | 3.36 | 6.77 | 6.78 | 4.75 | 4.64 | 2.93 | 3.41 | 3.18 | 63.0 | 66.6 | 82.1 | 82.4 | 74.7 | 73.8 |
| 1989 | 2.56 | 2.86 | 5.87 | 5.77 | 4.23 | 5.83 | 2.77 | 2.99 | 2.87 | 62.3 | 64.5 | 80.8 | 81.0 | 73.8 | 82.2 |
| 1990 | 2.53 | 2.61 | 6.47 | 5.78 | 3.90 | 5.09 | 2.67 | 2.72 | 2.69 | 62.3 | 62.7 | 83.4 | 81.1 | 72.6 | 78.6 |
| 1991 | 2.42 | 2.54 | 5.82 | 6.23 | 5.15 | 5.09 | 2.57 | 2.79 | 2.65 | 61.6 | 62.7 | 80.6 | 82.2 | 81.7 | 80.0 |
| 1992 | 2.54 | 2.66 | 6.49 | 6.01 | 4.09 | 5.28 | 2.86 | 2.74 | 2.81 | 62.3 | 63.2 | 83.4 | 81.1 | 77.4 | 82.7 |
| 1995 | 2.37 | 2.67 | 6.09 | 5.88 | 3.71 | 4.98 | 2.45 | 2.75 | 2.56 | 61.0 | 63.2 | 81.3 | 81.0 | 70.9 | 81.3 |
| 1996 | 2.63 | 2.86 | 6.50 | 6.30 | 4.98 | 5.44 | 2.83 | 2.90 | 2.88 | 62.8 | 64.0 | 81.4 | 81.1 | 77.1 | 79.4 |
| 1997 | 2.57 | 2.82 | 7.95 | 6.11 | 4.82 | 6.9 | 2.63 | 2.84 | 2.71 | 62.3 | 63.6 | 85.7 | 84.0 | 79.4 | 87.0 |
| 1998 | 2.72 | 2.83 | 6.44 | - | 3.28 | 4.77 | 2.76 | 2.84 | 2.78 | 62.0 | 62.7 | 84.0 | - | 66.3 | 76.0 |
| 1999 | 3.02 | 3.03 | 7.59 | - | 4.20 | - | 3.09 | 3.03 | 3.08 | 63.8 | 63.5 | 86.6 | - | 70.9 | - |
| 2000 | 2.47 | 2.81 | - | - | 2.58 | - | 2.47 | 2.81 | 2.57 | 60.7 | 63.2 | - | - | 64.7 | - |
| 2001 | 2.89 | 3.03 | 6.76 | 5.96 | 4.41 | 4.06 | 2.95 | 3.09 | 3.00 | 63.1 | 63.7 | 81.7 | 79.1 | 75.3 | 72.1 |
| 2002 | 2.84 | 2.92 | 7.12 | - | 5.00 | - | 2.89 | 2.92 | 2.90 | 62.6 | 62.1 | 83.0 | - | 75.8 | - |
| 2003 | 2.94 | 3.08 | 8.82 | 5.58 | 4.04 | - | 3.02 | 3.10 | 3.04 | 63 | 64.4 | 86.1 | 78.3 | 71.4 | - |
| 2004 | 3.11 | 2.95 | 7.33 | 5.22 | 4.71 | 6.48 | 3.17 | 3.22 | 3.18 | 64.7 | 65.0 | 86.2 | 76.4 | 77.6 | 88.0 |
| 2005 | 3.19 | 3.33 | 7.05 | 4.19 | 4.31 | 2.89 | 3.31 | 3.33 | 3.31 | 65.9 | 66.4 | 83.3 | 75.5 | 73.7 | 62.3 |
| 2006 | 3.10 | 3.25 | 9.72 |  | 5.05 | 3.67 | 3.25 | 3.26 | 3.24 | 65.3 | 65.3 | 90.0 |  | 76.8 | 69.5 |
| 2007 | 2.89 | 2.87 | 6.19 | 6.47 | 4.94 | 3.57 | 2.98 | 2.99 | 2.98 | 63.5 | 63.3 | 80.9 | 80.6 | 76.7 | 71.3 |
| 2008 | 3.04 | 3.03 | 6.35 | 7.47 | 3.82 | 3.39 | 3.08 | 3.07 | 3.08 | 64.6 | 63.9 | 80.1 | 85.5 | 71.1 | 73.0 |
| 2009 | 3.28 | 3.40 | 7.59 | 6.54 | 5.25 | 4.28 | 3.48 | 3.67 | 3.50 | 64.9 | 65.5 | 84.6 | 81.7 | 75.9 | 73.5 |
| 2010 | 3.44 | 3.24 | 6.40 | 5.45 | 4.17 | 3.92 | 3.47 | 3.28 | 3.42 | 66.7 | 65.2 | 80.0 | 75.0 | 72.4 | 70.0 |


|  | Whole weight (kg) |  |  |  |  |  | Fork Length (cm) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW |  | 2SW |  | PS |  | All se | ages | Total | 1 SW |  | 2SW |  | PS |  |
|  | NA | E | NA | E | NA | E | NA | E |  | NA | E | NA | E | NA | E |
| 2011 | 3.30 | 3.18 | 5.69 | 4.94 | 4.46 | 5.11 | 3.39 | 3.49 | 3.40 | 65.8 | 64.7 | 78.6 | 75.0 | 73.7 | 76.3 |
| 2012 | 3.34 | 3.38 | 6.00 | 4.51 | 4.65 | 3.65 | 3.44 | 3.40 | 3.44 | 65.5 | 64.9 | 75.9 | 70.4 | 72.8 | 68.9 |
| $\begin{aligned} & 10-\mathrm{yr} \\ & \text { mean } \end{aligned}$ | 3.16 | 3.17 | 7.11 | 5.60 | 4.54 | 4.11 | 3.26 | 3.28 | 3.26 | 65.0 | 64.9 | 82.6 | 77.6 | 74.2 | 72.5 |
| Overall mean | 2.85 | 3.17 | 6.61 | 6.24 | 4.10 | 4.70 | 3.00 | 3.25 | 3.12 | 63.3 | 65.3 | 81.9 | 81.2 | 71.8 | 75.9 |

Table 5.1.2.4. River age distribution (\%) and mean river age for all North American origin salmon caught at West Greenland 1968 to 1992 and 1995 to present. Continent of origin assignments were based on scale characteristics until 1995, scale characteristics and DNA based assignments until 2001 and DNA based assignments only from 2001 on.

| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 0.3 | 19.6 | 40.4 | 21.3 | 16.2 | 2.2 | 0 | 0 |
| 1969 | 0 | 27.1 | 45.8 | 19.6 | 6.5 | 0.9 | 0 | 0 |
| 1970 | 0 | 58.1 | 25.6 | 11.6 | 2.3 | 2.3 | 0 | 0 |
| 1971 | 1.2 | 32.9 | 36.5 | 16.5 | 9.4 | 3.5 | 0 | 0 |
| 1972 | 0.8 | 31.9 | 51.4 | 10.6 | 3.9 | 1.2 | 0.4 | 0 |
| 1973 | 2.0 | 40.8 | 34.7 | 18.4 | 2.0 | 2.0 | 0 | 0 |
| 1974 | 0.9 | 36 | 36.6 | 12.0 | 11.7 | 2.6 | 0.3 | 0 |
| 1975 | 0.4 | 17.3 | 47.6 | 24.4 | 6.2 | 4.0 | 0 | 0 |
| 1976 | 0.7 | 42.6 | 30.6 | 14.6 | 10.9 | 0.4 | 0.4 | 0 |
| 1977 | - | - | - | - | - | - | - | - |
| 1978 | 2.7 | 31.9 | 43.0 | 13.6 | 6.0 | 2.0 | 0.9 | 0 |
| 1979 | 4.2 | 39.9 | 40.6 | 11.3 | 2.8 | 1.1 | 0.1 | 0 |
| 1980 | 5.9 | 36.3 | 32.9 | 16.3 | 7.9 | 0.7 | 0.1 | 0 |
| 1981 | 3.5 | 31.6 | 37.5 | 19.0 | 6.6 | 1.6 | 0.2 | 0 |
| 1982 | 1.4 | 37.7 | 38.3 | 15.9 | 5.8 | 0.7 | 0 | 0.2 |
| 1983 | 3.1 | 47.0 | 32.6 | 12.7 | 3.7 | 0.8 | 0.1 | 0 |
| 1984 | 4.8 | 51.7 | 28.9 | 9.0 | 4.6 | 0.9 | 0.2 | 0 |
| 1985 | 5.1 | 41.0 | 35.7 | 12.1 | 4.9 | 1.1 | 0.1 | 0 |
| 1986 | 2.0 | 39.9 | 33.4 | 20.0 | 4.0 | 0.7 | 0 | 0 |
| 1987 | 3.9 | 41.4 | 31.8 | 16.7 | 5.8 | 0.4 | 0 | 0 |
| 1988 | 5.2 | 31.3 | 30.8 | 20.9 | 10.7 | 1.0 | 0.1 | 0 |
| 1989 | 7.9 | 39.0 | 30.1 | 15.9 | 5.9 | 1.3 | 0 | 0 |
| 1990 | 8.8 | 45.3 | 30.7 | 12.1 | 2.4 | 0.5 | 0.1 | 0 |
| 1991 | 5.2 | 33.6 | 43.5 | 12.8 | 3.9 | 0.8 | 0.3 | 0 |
| 1992 | 6.7 | 36.7 | 34.1 | 19.1 | 3.2 | 0.3 | 0 | 0 |
| 1993 | - | - | - | - | - | - | - | - |
| 1994 | - | - | - | - | - | - | - | - |
| 1995 | 2.4 | 19.0 | 45.4 | 22.6 | 8.8 | 1.8 | 0.1 | 0 |
| 1996 | 1.7 | 18.7 | 46.0 | 23.8 | 8.8 | 0.8 | 0.1 | 0 |
| 1997 | 1.3 | 16.4 | 48.4 | 17.6 | 15.1 | 1.3 | 0 | 0 |
| 1998 | 4.0 | 35.1 | 37.0 | 16.5 | 6.1 | 1.1 | 0.1 | 0 |
| 1999 | 2.7 | 23.5 | 50.6 | 20.3 | 2.9 | 0.0 | 0 | 0 |
| 2000 | 3.2 | 26.6 | 38.6 | 23.4 | 7.6 | 0.6 | 0 | 0 |
| 2001 | 1.9 | 15.2 | 39.4 | 32.0 | 10.8 | 0.7 | 0 | 0 |
| 2002 | 1.5 | 27.4 | 46.5 | 14.2 | 9.5 | 0.9 | 0 | 0 |
| 2003 | 2.6 | 28.8 | 38.9 | 21.0 | 7.6 | 1.1 | 0 | 0 |
| 2004 | 1.9 | 19.1 | 51.9 | 22.9 | 3.7 | 0.5 | 0 | 0 |
| 2005 | 2.7 | 21.4 | 36.3 | 30.5 | 8.5 | 0.5 | 0 | 0 |
| 2006 | 0.6 | 13.9 | 44.6 | 27.6 | 12.3 | 1.0 | 0 | 0 |
| 2007 | 1.6 | 27.7 | 34.5 | 26.2 | 9.2 | 0.9 | 0 | 0 |


| YEAR | $\mathbf{1}$ | $\mathbf{2}$ |  | $\mathbf{3}$ |  | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 0.9 | 25.1 | 51.9 | 16.8 | 4.7 | 0.6 | 0 | $\mathbf{8}$ |  |
| 2009 | 2.6 | 30.7 | 47.3 | 15.4 | 3.7 | 0.4 | 0 | 0 |  |
| 2010 | 1.6 | 21.7 | 47.9 | 21.7 | 6.3 | 0.8 | 0 | 0 |  |
| 2011 | 1.0 | 35.9 | 45.9 | 14.4 | 2.8 | 0 | 0 | 0 |  |
| 2012 | 0.3 | 29.8 | 39.4 | 23.3 | 6.5 | 0.7 | 0 | 0 |  |
| $10-y r$ mean | 1.6 | 25.4 | 43.9 | 22.0 | 6.5 | 0.7 | 0.0 | 0.0 |  |
| Overall Mean | 2.6 | 31.6 | 39.6 | 18.3 | 6.7 | 1.1 | 0.1 | 0.0 |  |

Table 5.1.2.5. River age distribution (\%) and mean river age for all European origin salmon caught at West Greenland 1968 to 1992 and 1995 to present. Continent of origin assignments were based on scale characteristics until 1995, scale characteristics and DNA based assignments until 2001 and DNA based assignments only from 2001 on.

| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 21.6 | 60.3 | 15.2 | 2.7 | 0.3 | 0 | 0 | 0 |
| 1969 | 0 | 83.8 | 16.2 | 0 | 0 | 0 | 0 | 0 |
| 1970 | 0 | 90.4 | 9.6 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 9.3 | 66.5 | 19.9 | 3.1 | 1.2 | 0 | 0 | 0 |
| 1972 | 11.0 | 71.2 | 16.7 | 1.0 | 0.1 | 0 | 0 | 0 |
| 1973 | 26.0 | 58.0 | 14.0 | 2.0 | 0 | 0 | 0 | 0 |
| 1974 | 22.9 | 68.2 | 8.5 | 0.4 | 0 | 0 | 0 | 0 |
| 1975 | 26.0 | 53.4 | 18.2 | 2.5 | 0 | 0 | 0 | 0 |
| 1976 | 23.5 | 67.2 | 8.4 | 0.6 | 0.3 | 0 | 0 | 0 |
| 1977 | - | - | - | - | - | - | - | - |
| 1978 | 26.2 | 65.4 | 8.2 | 0.2 | 0 | 0 | 0 | 0 |
| 1979 | 23.6 | 64.8 | 11.0 | 0.6 | 0 | 0 | 0 | 0 |
| 1980 | 25.8 | 56.9 | 14.7 | 2.5 | 0.2 | 0 | 0 | 0 |
| 1981 | 15.4 | 67.3 | 15.7 | 1.6 | 0 | 0 | 0 | 0 |
| 1982 | 15.6 | 56.1 | 23.5 | 4.2 | 0.7 | 0 | 0 | 0 |
| 1983 | 34.7 | 50.2 | 12.3 | 2.4 | 0.3 | 0.1 | 0.1 | 0 |
| 1984 | 22.7 | 56.9 | 15.2 | 4.2 | 0.9 | 0.2 | 0 | 0 |
| 1985 | 20.2 | 61.6 | 14.9 | 2.7 | 0.6 | 0 | 0 | 0 |
| 1986 | 19.5 | 62.5 | 15.1 | 2.7 | 0.2 | 0 | 0 | 0 |
| 1987 | 19.2 | 62.5 | 14.8 | 3.3 | 0.3 | 0 | 0 | 0 |
| 1988 | 18.4 | 61.6 | 17.3 | 2.3 | 0.5 | 0 | 0 | 0 |
| 1989 | 18.0 | 61.7 | 17.4 | 2.7 | 0.3 | 0 | 0 | 0 |
| 1990 | 15.9 | 56.3 | 23.0 | 4.4 | 0.2 | 0.2 | 0 | 0 |
| 1991 | 20.9 | 47.4 | 26.3 | 4.2 | 1.2 | 0 | 0 | 0 |
| 1992 | 11.8 | 38.2 | 42.8 | 6.5 | 0.6 | 0 | 0 | 0 |
| 1993 | - | - | - | - | - | - | - | - |
| 1994 | - | - | - | - | - | - | - | - |
| 1995 | 14.8 | 67.3 | 17.2 | 0.6 | 0 | 0 | 0 | 0 |
| 1996 | 15.8 | 71.1 | 12.2 | 0.9 | 0 | 0 | 0 | 0 |
| 1997 | 4.1 | 58.1 | 37.8 | 0.0 | 0 | 0 | 0 | 0 |
| 1998 | 28.6 | 60.0 | 7.6 | 2.9 | 0.0 | 1.0 | 0 | 0 |
| 1999 | 27.7 | 65.1 | 7.2 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 36.5 | 46.7 | 13.1 | 2.9 | 0.7 | 0 | 0 | 0 |
| 2001 | 16.0 | 51.2 | 27.3 | 4.9 | 0.7 | 0 | 0 | 0 |
| 2002 | 9.4 | 62.9 | 20.1 | 7.6 | 0 | 0 | 0 | 0 |
| 2003 | 16.2 | 58.0 | 22.1 | 3.0 | 0.8 | 0 | 0 | 0 |
| 2004 | 18.3 | 57.7 | 20.5 | 3.2 | 0.2 | 0 | 0 | 0 |
| 2005 | 19.2 | 60.5 | 15.0 | 5.4 | 0 | 0 | 0 | 0 |
| 2006 | 17.7 | 54.0 | 23.6 | 3.7 | 0.9 | 0 | 0 | 0 |
| 2007 | 7.0 | 48.5 | 33.0 | 10.5 | 1.0 | 0 | 0 | 0 |
| 2008 | 7.0 | 72.8 | 19.3 | 0.8 | 0.0 | 0 | 0 | 0 |
| 2009 | 14.3 | 59.5 | 23.8 | 2.4 | 0.0 | 0 | 0 | 0 |
| 2010 | 11.3 | 57.1 | 27.3 | 3.4 | 0.8 | 0 | 0 | 0 |
| 2011 | 18.3 | 54.9 | 25.4 | 1.4 | 0 | 0 | 0 | 0 |
| 2012 | 9.3 | 63.0 | 24.0 | 3.7 | 0 | 0 | 0 | 0 |
| 10-yr mean | 13.9 | 58.6 | 23.4 | 3.7 | 0.4 | 0.0 | 0.0 | 0.0 |
| Overall Mean | 17.6 | 60.9 | 18.5 | 2.7 | 0.3 | 0.0 | 0.0 | 0.0 |

Table 5.1.2.6. Sea age composition (\%) of samples from fishery landings at West Greenland from 1985 by continent of origin.

|  | North American |  |  | European |  |  |
| :---: | ---: | ---: | :---: | ---: | :---: | :---: |
| Year | 1SW | 2SW | Previous <br> Spawners | 1SW | 2SW | Previous <br> Spawners |
| 1985 | 92.5 | 7.2 | 0.3 | 95.0 | 4.7 | 0.4 |
| 1986 | 95.1 | 3.9 | 1.0 | 97.5 | 1.9 | 0.6 |
| 1987 | 96.3 | 2.3 | 1.4 | 98.0 | 1.7 | 0.3 |
| 1988 | 96.7 | 2.0 | 1.2 | 98.1 | 1.3 | 0.5 |
| 1989 | 92.3 | 5.2 | 2.4 | 95.5 | 3.8 | 0.6 |
| 1990 | 95.7 | 3.4 | 0.9 | 96.3 | 3.0 | 0.7 |
| 1991 | 95.6 | 4.1 | 0.4 | 93.4 | 6.5 | 0.2 |
| 1992 | 91.9 | 8.0 | 0.1 | 97.5 | 2.1 | 0.4 |
| 1993 | - | - | - | - | - | - |
| 1994 | - | - | - | - | - | - |
| 1995 | 96.8 | 1.5 | 1.7 | 97.3 | 2.2 | 0.5 |
| 1996 | 94.1 | 3.8 | 2.1 | 96.1 | 2.7 | 1.2 |
| 1997 | 98.2 | 0.6 | 1.2 | 99.3 | 0.4 | 0.4 |
| 1998 | 96.8 | 0.5 | 2.7 | 99.4 | 0.0 | 0.6 |
| 1999 | 96.8 | 1.2 | 2.0 | 100.0 | 0.0 | 0.0 |
| 2000 | 97.4 | 0.0 | 2.6 | 100.0 | 0.0 | 0.0 |
| 2001 | 98.2 | 2.6 | 0.5 | 97.8 | 2.0 | 0.3 |
| 2002 | 97.3 | 0.9 | 1.8 | 100.0 | 0.0 | 0.0 |
| 2003 | 96.7 | 1.0 | 2.3 | 98.9 | 1.1 | 0.0 |
| 2004 | 97.0 | 0.5 | 2.5 | 97.0 | 2.8 | 0.2 |
| 2005 | 92.4 | 1.2 | 6.4 | 96.7 | 1.1 | 2.2 |
| 2006 | 93.0 | 0.8 | 5.6 | 98.8 | 0.0 | 1.2 |
| 2007 | 96.5 | 1.0 | 2.5 | 95.6 | 2.5 | 1.5 |
| 2008 | 97.4 | 0.5 | 2.2 | 98.8 | 0.8 | 0.4 |
| 2009 | 93.4 | 2.8 | 3.8 | 89.4 | 7.6 | 3.0 |
| 2010 | 98.2 | 0.4 | 1.4 | 97.5 | 1.7 | 0.8 |
| 2011 | 93.8 | 1.5 | 4.7 | 82.8 | 12.1 | 5.2 |
| 2012 | 93.2 | 0.7 | 6.0 | 98.0 | 1.6 | 0.4 |
|  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |

Table 5.1.3.1. The number of samples and continent of origin of Atlantic salmon by NAFO Division sampled at West Greenland in 2012. NA = North America, E = Europe.

|  |  | Numbers |  |  | Percentaces |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| NAFO Div | Sample dates | NA | E | Totals | NA | E |
|  |  |  |  |  |  |  |
| 1B | September 3-October 1 | 442 | 22 | 464 | 95.3 | 4.7 |
|  |  |  |  |  |  |  |
| 1C | September 21-October 7 | 431 | 154 | 585 | 73.7 | 26.3 |
|  |  |  |  |  |  |  |
| 1F |  | 248 | 76 | 324 | 76.5 | 23.5 |
|  |  | 1121 | 252 | 1373 | 81.6 | 18.4 |
| Total |  |  |  |  |  |  |

Table 5.1.3.2. The numbers of North American (NA) and European (E) Atlantic salmon caught at West Greenland 1971 to 1992 and 1995 to present and the proportion by continent of origin, based on NAFO Division continent of origin weighted by catch (weight) in each division. Numbers are rounded to the nearest hundred fish.

|  | Proportion by continent weighted by catch in number |  | Numbers of salmon by continent |  |
| :---: | :---: | :---: | :---: | :---: |
|  | NA | E | NA | E |
| 1982 | 57 | 43 | 192200 | 143800 |
| 1983 | 40 | 60 | 39500 | 60500 |
| 1984 | 54 | 46 | 48800 | 41200 |
| 1985 | 47 | 53 | 143500 | 161500 |
| 1986 | 59 | 41 | 188300 | 131900 |
| 1987 | 59 | 41 | 171900 | 126400 |
| 1988 | 43 | 57 | 125500 | 168800 |
| 1989 | 55 | 45 | 65000 | 52700 |
| 1990 | 74 | 26 | 62400 | 21700 |
| 1991 | 63 | 37 | 111700 | 65400 |
| 1992 | 45 | 55 | 46900 | 38500 |
| 1995 | 67 | 33 | 21400 | 10700 |
| 1996 | 70 | 30 | 22400 | 9700 |
| 1997 | 85 | 15 | 18000 | 3300 |
| 1998 | 79 | 21 | 3100 | 900 |
| 1999 | 91 | 9 | 5700 | 600 |
| 2000 | 65 | 35 | 5100 | 2700 |
| 2001 | 67 | 33 | 9400 | 4700 |
| 2002 | 69 | 31 | 2300 | 1000 |
| 2003 | 64 | 36 | 2600 | 1400 |
| 2004 | 72 | 28 | 3900 | 1500 |
| 2005 | 74 | 26 | 3500 | 1200 |
| 2006 | 69 | 31 | 4000 | 1800 |
| 2007 | 76 | 24 | 6100 | 1900 |
| 2008 | 86 | 14 | 8000 | 1300 |
| 2009 | 90 | 10 | 7000 | 800 |
| 2010 | 81 | 19 | 10000 | 2600 |
| 2011 | 91 | 9 | 6800 | 600 |
| 2012 | 79 | 21 | 7800 | 2100 |



Figure 5.1.1.1. Exploitation rate (\%) for NAC 1SW non-maturing and southern NEAC nonmaturing Atlantic salmon at West Greenland, 1971-2011. Exploitation rate estimates are only available to 2011, as 2012 exploitation rates are dependent on 2013 2SW NAC or MSW (NEAC) returns.


Figure 5.1.1.2. Nominal catches and commercial quotas (metric tonnes, round fresh weight) of salmon at West Greenland for 1971-2012 (top panel) and 2003-2012 (bottom panel). The quota has been set to nil since 2003.


Figure 5.1.1.3. Location of NAFO divisions along the coast of West Greenland. Stars identify the communities where biological sampling occurred (Sisimiut, Maniitsoq and Qaqortoq).


Figure 5.1.3.1. Percent of the sampled catch by continent of origin for the 1982 to 2012 Atlantic salmon West Greenland fishery.


Figure 5.1.3.2. Number of North American and European Atlantic salmon caught at West Greenland from 1982 to 2012 (upper panel) and 2003 to 2012 (lower panel).

## Annex 1: Working documents submitted to the Working Group on North Atlantic Salmon, 3-12 April, 2013

| WP No. | Authors | Title |
| :---: | :---: | :---: |
| 1 | Trial, J., Sweka, J., Kocik, J., Sheehan, T., Freidland, K. and Letcher, B. | National Report for the United States, 2012. |
| 2 | Sheehan, T. F., Assunção, M. G. L., <br> Deschamps, D., Laughton, B., Ó Cuaig, M., Nygaard, R., King, T. L., Robertson, M. J. and $\mathrm{O}^{\prime}$ Maoiléidigh, N . | The International Sampling Programme: <br> Continent of Origin and Biological Characteristics of Atlantic Salmon Collected at West Greenland in 2012. |
| 3 | Sheehan, T. F., Assunção, M. G. L., Chisholm, N., Deschamps, D., Dixon, H., Renkawitz, M. Rogan, G., Nygaard, R., King, T. L., Robertson, M. J. and O' Maoiléidigh, N . | The International Sampling Programme: Update of the Continent of Origin and Biological Characteristics of Atlantic Salmon Collected at West Greenland in 2011. |
| 4 | Nieland, J. L., Sheehan, T. F., Saunders, R., Murphy, J. S., Trinko Lake, T. and Stevens, J. R. | Dam Impact Analysis Model for Atlantic Salmon in the Penobscot River, Maine. |
| 5 | Mills, K. E., Pershing, A. J., Sheehan, T. F. and Mountain, D. | Climate and ecosystem linkages explain the widespread decline in North American Atlantic salmon populations. |
| 6 | Renkawitz, M. D. and Sheehan T. F. | West Greenland foraging ecology and implications for survival. |
| 7 | Renkawitz, M. D., Sheehan, T. F., Rikardsen, A., Chittenden, C. Nygaard, R. and Righton, D. | Tagging adult Atlantic salmon at West Greenland with pop-up archival satellite tags. |
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| 9 | Meerburg, D. | Atlantic Salmon Federation tracking studies |
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|  |  | Meeting WGNAS. |
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| 23 | Degerman, E., Persson, J., Sers, B. and Östergren, J. | Fisheries management and status of Atlantic salmon stocks in Sweden: National Report for 2012. |
| 24 | Rasmussen, G. | Catch data for Denmark for 2012. |
| 25 | Nygaard, R. | The salmon fishery in Greenland, 2012. |
| 26 | White, J. | ECOKNOWS update. |
| 27 | Chaput, Potter, Saunders and Gauldbek | NASCO West Greenland Commission Report of the Framework of Indicators Working Group 2013 . |
| 28 | Russell, I.C., Fiske, P., Prusov, S. and Jacobsen, J-A. | NASCO North East Atlantic Commission Report of the Framework of Indicators Working Group 2013. |
| 29 | Levy, A. L., R. A. Jones, S. C. Hansen and A. J. F. Gibson | Status of Atlantic salmon in Canada's MaritImes region (Salmon Fishing Areas 19 to 23) |
| 30 | Robertson, M., Poole, R., Levy, A., Jones, R., Douglas, S., Dionne, M., Cameron, P., Cairns, D., Breau, C., and Chaput, G. | Catch Statistics and Aquaculture <br> Production Values for Canada: preliminary 2012 and final 2011. |
| 31 | Breau, C., Cameron, P., Douglas, S., and Chaput, G. | Stock status update for Gulf Region Salmon Fishing Areas 15 to 18. |
| 32 | Chaput, G. | Monitoring of sea lice burdens on wild returning adult Atlantic salmon from the Miramichi River, New Brunswick. |
| 33 | Chaput, G., Carr, J., Jonsen, I., and Whoriskey, F. | Modelling inter-stage survival rates and dectection probabilities for acoustically tracked Atlantic salmon smolts and postsmolts: model, assumptions, diagnostics, considerations for planning experiments. |
| 34 | Dionne, M. | Stock status of Atlantic salmon in Quebec for 2012. |
| 35 | Dionne, M. | Index rivers monitoring programme for Quebec to 2012. |
| 36 | Massiot-Granier, F., Prévost, E., Chaput, G., Potter, E.C.E., Smith, G., White, J., Mäntyniemi, S. and Rivot, E. | Embedding Atlantic salmon stock assessment at broad ocean scale within an integrated Bayesian life cycle modelling framework. |
| 37 | "White, J., Ó Maoiléidigh, N., Gargan, P., De Eyto, E., McGinnity, P., Roche, W., Lawlor, I., Kennedy, B. and, Doherty, D. | Updating River Conservation Limits for Irish Salmon Rivers. |
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## Annex 3: Participants list

| Name | Address | Phone/Fax | E-MAIL |
| :---: | :---: | :---: | :---: |
| Gérald Chaput | Fisheries and Oceans Canada DFO Moncton PO Box 5030 <br> Moncton NB E1C 9B6 Canada | $\begin{aligned} & \text { Phone +1 } 506 \\ & 8512022 \\ & \text { Fax }+1506851 \\ & 2620 \end{aligned}$ | Gerald.Chaput@dfo-mpo.gc.ca |
| Dennis Ensing | Agri-food and Biosciences Institute (AFBI) <br> Fisheries \& Aquatic Ecosystems Branch Newforge Lane BT9 5PX Belfast Northern Ireland United Kingdom | $\begin{aligned} & \text { Phone }+442890 \\ & 255054 \\ & \text { Fax +44 } 28255 \\ & 004 \end{aligned}$ | dennis.ensing@afbini.gov.uk |
| Jaakko <br> Erkinaro | Finnish Game and Fisheries Research Institute <br> Oulu Game and Fisheries Research PO 413 <br> 90014 Oulu <br> Finland | $\begin{aligned} & \text { Phone +358 } 295 \\ & 327871 \\ & \text { Fax +358 } 20575 \\ & 1879 \end{aligned}$ | jaakko.erkinaro@rktl.fi |
| Gilles Euzenat | ONEMA, Dast <br> Station d'Ecologie <br> Piscicole <br> r. des Fontaines <br> 76260 EU <br> France | $\begin{aligned} & \text { Phone +33 } 227 \\ & 280611 \\ & \text { Fax +33 } 23582 \\ & 6207 \end{aligned}$ | gilles.euzenat@onema.fr |
| Peder Fiske | Norwegian Institute for Nature Research N-7485 Trondheim Norway | $\begin{aligned} & \text { Phone }+47 \\ & 93466733 \end{aligned}$ | Peder.Fiske@nina.no |
| Harald <br> Gjøsæter | Institute of Marine Research <br> PO Box 1870 <br> Nordnes <br> 5817 Bergen <br> Norway | Phone +47 <br> 55238417 / mob +4741479177 <br> Fax +4755 <br> 238687 | Harald.Gjoesaeter@imr.no |
| Gudni Gudbergsson | Institute of Freshwater Fisheries <br> 'Arleyni 22 <br> IS-112 Reykjavik <br> Iceland | $\begin{aligned} & \text { Phone }+354 \\ & 5806300 \\ & \text { Fax }+354 \\ & 5806301 \end{aligned}$ | gudni.gudbergsson@veidimal.is |


| Name | Address | Phone/Fax | E-MAIL |
| :---: | :---: | :---: | :---: |
| Félix MassiotGranier | Agrocampus Ouest UMR INRA- <br> Agrocampus Ecology et Santé des Ecosysteme 65, rue de St Brieuc 35045 Rennes France | Phone +33 <br> 617754787 <br> Fax +33 | felix.massiotgranier@gmail.com |
| Dave Meerburg | Atlantic Salmon <br> Federation <br> PO Box 5200 <br> St Andrews NB E5B 3S8 <br> Canada | Phone +1 613 <br> 9900286 <br> Fax +1 613954 0807 | dmeerburg@asf.ca |
| Rasmus <br> Nygaard <br> By <br> correspondence | Greenland Institute for Natural Resources <br> PO Box 570 <br> 3900 Nuuk <br> Greenland | $\begin{aligned} & \text { Phone }+299 \\ & \text { Fax }+299 \end{aligned}$ | RaNY@natur.gl |
| Niall OMMaoiléidigh | Marine Institute <br> Fisheries Ecosystem <br> Advisory Services <br> Farran Laboratory <br> Furnace, Newport <br> Co. Mayo <br> Ireland | Phone +353 <br> 9842300 <br> Fax +353 <br> 9842340 | niall.omaoileidigh@marine.ie |
| James <br> Orpwood | Marine Scotland <br> Science <br> Freshwater Laboratory <br> Faskally, Pitlochry <br> Perthshire PH16 5LB <br> United Kingdom | $\begin{aligned} & \text { Phone }+441796 \\ & 472060 \\ & \text { Fax }+441796 \\ & 473523 \end{aligned}$ | James.orpwood@scotland.gsi.gov.uk |
| Ted Potter | Centre for <br> Environment, Fisheries and Aquaculture Science (Cefas) Lowestoft Laboratory Pakefield Road NR33 0HT Lowestoft Suffolk United Kingdom | Phone +44 1502 <br> 524560 <br> Fax +441502 <br> 513865 | ted.potter@cefas.co.uk |
| Sergey Prusov | Knipovich Polar Research Institute of Marine Fisheries and Oceanography(PINRO) 6 Knipovitch Street 183038 Murmansk Russian Federation | $\begin{aligned} & \text { Phone }+78152 \\ & 473658 \\ & \text { Fax }+78152 \\ & 473331 \end{aligned}$ | prusov@pinro.ru |


| Name | Address | Phone/Fax | E-MAIL |
| :---: | :---: | :---: | :---: |
| Etienne Rivot | Agrocampus Ouest UMR INRA- <br> Agrocampus Ecology et Santé des Ecosysteme 65, rue de St Brieuc 35045 Rennes France | $\begin{aligned} & \text { Phone +33 } 223 \\ & 485934 \end{aligned}$ | etienne.rivot@agrocampus-ouest.fr |
| Martha <br> Robertson | Fisheries and Oceans <br> Canada <br> Northwest Atlantic <br> Fisheries Center <br> PO Box 5667 <br> St John's NL A1C 5X1 <br> Canada | $\begin{aligned} & \text { Phone }+1709 \\ & 7724553 \\ & \text { Fax }+1 \end{aligned}$ | martha.robertson@dfo-mpo.gc.ca |
| Ian Russell <br> Chair | Centre for Environment, Fisheries and Aquaculture Science (Cefas) Lowestoft Laboratory Pakefield Road NR33 0HT Lowestoft Suffolk United Kingdom | $\begin{aligned} & \text { Phone }+441502 \\ & 524330 \\ & \text { Fax +44 } 1502 \\ & 513865 \end{aligned}$ | ian.russell@cefas.co.uk |
| Tim Sheehan | National Marine <br> Fisheries Services <br> Northeast Fisheries <br> Science Center <br> Woods Hole <br> Laboratory <br> 166 Water Street <br> Woods Hole MA 02543 <br> United States | Phone +1 508495-2215 <br> Fax +1 508495- $2393$ | tim.sheehan@noaa.gov |
| Gordon Smith | Marine Scotland <br> Science <br> Freshwater Laboratory <br> Field Station <br> Inchbraoch House, <br> South Quay <br> DD10 9SL Ferryden <br> Montrose Angus <br> United Kingdom | $\begin{aligned} & \text { Phone + } 441674 \\ & 677070 \\ & \text { Fax + 44 } 1674 \\ & 672604 \end{aligned}$ | gordon.smith@scotland.gsi.gov.uk |
| Gennady <br> Ustyuzhinsky | Knipovich Polar <br> Research Institute of <br> Marine Fisheries and <br> Oceanography(PINRO) <br> 17, Uritskogo Street <br> RU-163002 <br> Arkhangelsk <br> Russian Federation | Phone +78182 <br> 661646 <br> Fax +78182 <br> 661650 | gena@pinro.ru |


| Name | AdDRESS | Phone/FAX | E-MAIL |
| :--- | :--- | :--- | :--- |
| Jonathan White | Marine Institute | Phone +35391 | jonathan.white@marine.ie |
|  | Rinville | 387200 |  |
|  | Oranmore | Fax +353 |  |
|  | Co. Galway | 91387201 |  |
|  | Ireland |  |  |

## Annex 4: Reported catch of salmon by sea age class

Reported catch of salmon in numbers and weight (tonnes round fresh weight) by sea age class. Catches reported for 2012 may be provisional. Methods used for estimating age composition given in footnote.

West Greenland


## Canada



USA


Faroe Islands


Finland

| Country | Year | 1SW |  | 2SW |  | 3SW |  | 4SW |  | 5SW |  | MSW (1) |  | PS |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt |
| Finland | 1982 | 2598 | 5 | - | - |  |  | - |  | - |  | 5408 | 49 |  |  | 8006 | 54 |
|  | 1983 | 3916 | 7 | - | - |  | - | - |  |  |  | 6050 | 51 | - | - | 9966 | 58 |
|  | 1984 | 4899 | 9 | - | - | - | - | - |  | - |  | 4726 | 37 | - | - | 9625 | 46 |
|  | 1985 | 6201 | 11 | - | - | - | - | - |  | - |  | 4912 | 38 | - | - | 11113 | 49 |
|  | 1986 | 6131 | 12 | - | - |  | - | - |  | - |  | 3244 | 25 | - | - | 9375 | 37 |
|  | 1987 | 8696 | 15 | - | - | - | - | - |  | - |  | 4520 | 34 | - | - | 13216 | 49 |
|  | 1988 | 5926 | 9 | - | - |  | - | - |  | - |  | 3495 | 27 | - | - | 9421 | 36 |
|  | 1989 | 10395 | 19 | - | - | - | - | - |  | - |  | 5332 | 33 |  | - | 15727 | 52 |
|  | 1990 | 10084 | 19 |  | - |  | - | - |  | - |  | 5600 | 41 |  | - | 15684 | 60 |
|  | 1991 | 9213 | 17 |  | - |  | - | - |  | - |  | 6298 | 53 |  | - | 15511 | 70 |
|  | 1992 | 15017 | 28 |  | - |  | - | - |  | - |  | 6284 | 49 |  | - | 21301 | 77 |
|  | 1993 | 11157 | 17 | - | - | - | - | - |  | - |  | 8180 | 53 | - | - | 19337 | 70 |
|  | 1994 | 7493 | 11 | - | - |  | - | - |  | - |  | 6230 | 38 | - | - | 13723 | 49 |
|  | 1995 | 7786 | 11 | - | - |  | - | - |  |  |  | 5344 | 38 | - | - | 13130 | 49 |
|  | 1996 | 12230 | 20 | 1275 | 5 | 1424 | 12 | 234 |  | 19 |  |  | - | 354 | 3 | 15536 | 44 |
|  | 1997 | 10341 | 15 | 2419 | 10 | 1674 | 15 | 141 |  | 22 |  |  |  | 418 | 3 | 15015 | 45 |
|  | 1998 | 11792 | 19 | 1608 | 7 | 1660 | 16 | 147 |  | - |  |  |  | 460 | 3 | 15667 | 48 |
|  | 1999 | 18830 | 33 | 1528 | 8 | 1579 | 16 | 129 |  | 6 |  | - | - | 490 | 3 | 22562 | 62 |
|  | 2000 | 20817 | 39 | 5152 | 24 | 2379 | 25 | 110 |  |  |  | - | - | 991 | 6 | 56000 | 95 |
|  | 2001 | 13296 | 21 | 6286 | 32 | 5369 | 57 | 103 |  |  |  | - |  | 2372 | 13 | 27426 | 125 |
|  | 2002 | 6427 | 12 | 5227 | 20 | 4048 | 43 | 145 |  | 11 |  | - | - | 2496 | 16 | 18354 | 93 |
|  | 2003 | 8130 | 15 | 1828 | 7 | 3599 | 35 | 161 |  | 6 |  | - | - | 2204 | 15 | 15928 | 75 |
|  | 2004 | 3849 | 7 | 1425 | 6 | 1152 | 11 | 251 |  | 6 |  | - |  | 1404 | 11 | 8087 | 39 |
|  | 2005 | 9216 | 16 | 1027 | 5 | 1575 | 16 | 90 |  | 66 |  | 3595 | - | 837 | 8 | 12812 | 47 |
|  | 2006 | 17758 | 29 | 4166 | 18 | 1369 | 13 | 66 |  |  |  | 6370 |  | 770 | 5 | 24128 | 67 |
|  | 2007 | 3250 | 6 | 5329 | 21 | 2423 | 23 | 23 |  | 7 |  | 9038 |  | 1255 | 8 | 12288 | 59 |
|  | 2008 | 3347 | 6 | 1848 | 8 | 4394 | 41 | 235 |  |  |  | 8378 |  | 1901 | 11 | 11726 | 71 |
|  | 2009 | 6727 |  | 1531 | - | 1408 | - | 262 |  |  |  | 4077 |  | 876 | - | 10804 | 36 |
|  | 2010 | 6361 |  | 3687 | - | 1345 | - | 322 |  | 12 |  | 6322 |  | 957 | - | 12684 | 49 |
|  | 2011 | 7782 | - | 2078 | - | 1725 | - | 190 |  | 22 |  | 4926 | - | 911 | - | 12709 | 44 |
|  | 2012 | 15331 | 31 | 3108 | 12 | 1302 | 12 | 188 |  | 10 |  | 5592 | 33 | 984 | 6 | 20923 | 64 |

Iceland

| Iceland | 1991 | 29601 |  | 11892 |  |  |  |  |  |  |  |  |  |  |  | 41493 | 130 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1992 | 38538 |  | 15312 |  | - | - | - | - |  | - | - | - | - | - | 53850 | 175 |
|  | 1993 | 36640 |  | 11541 | - | - | - | - | - | - | - | - |  | - | - | 48181 | 160 |
|  | 1994 | 24224 | 59 | 14088 | 76 | - | - | - | - | - | - | - |  | - | - | 38312 | 135 |
|  | 1995 | 32767 | 90 | 13136 | 56 | - | - | - | - | - | - | - |  | - | - | 45903 | 145 |
|  | 1996 | 26927 | 66 | 9785 | 52 | - | - | - | - | - | - | - |  | - | - | 36712 | 118 |
|  | 1997 | 21684 | 56 | 8178 | 41 | - | - | - | - | - | - | - | - | - | - | 29862 | 97 |
|  | 1998 | 32224 | 81 | 7272 | 37 | - | - | - | - | - | - | - | - | - | - | 39496 | 119 |
|  | 1999 | 22620 | 59 | 9883 | 52 | - | - | - | - | - | - | - | - | - | - | 32503 | 111 |
|  | 2000 | 20270 | 49 | 4319 | 24 | - | - | - | - | - | - | - | - | - | - | 24589 | 73 |
|  | 2001 | 18538 | 46 | 5289 | 28 | - | - | - | - | - | - | - | - | - | - | 23827 | 74 |
|  | 2002 | 25277 | 64 | 5194 | 26 | - | - | - | - | - | - | - | - | - | - | 30471 | 90 |
|  | 2003 | 24738 | 61 | 8119 | 37 | - | - | - | - | - | - | - | - | - | - | 32857 | 99 |
|  | 2004 | 32600 | 84 | 6128 | 28 | - | - | - | - | - | - | - | - | - | - | 38728 | 111 |
|  | 2005 | 39980 | 101 | 5941 | 28 | - | - | - | - | - | - | - | - | - | - | 45921 | 129 |
|  | 2006 | 29857 | 71 | 5635 | 23 | - | - | - | - | - | - | - | - | - | - | 35492 | 93 |
|  | 2007 | 31899 | 74 | 3262 | 15 | - | - | - | - | - | - | - | - | - | - | 35161 | 89 |
|  | 2008 | 44391 | 106 | 5129 | 26 | - | - | - | - | - | - | - | - | - | - | 49520 | 132 |
|  | 2009 | 43981 | 103 | 4561 | 24 | - | - | - | - | - | - | - | - | - | - | 48542 | 126 |
|  | 2010 | 43457 | 105 | 9251 | 43 | - | - | - | - | - | - | - | - | - | - | 52708 | 147 |
|  | 2011 | 28550 | 74 | 4854 | 24 | - | - | - | - | - | - | - | - | - | - | 33404 | 98 |
|  | 2012 | 18033 | 39 | 2732 | 18 | - | - | - | - | - | - | - | - | - | - | 20765 | 57 |

Sweden


## Norway

| Country | Year | 1SW |  | 2SW |  | 3SW |  | 4SW |  | 5SW |  | MSW (1) |  | PS |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt |
| Norway | 1981 | 221566 | 467 |  |  |  |  |  |  |  |  | 213943 | 1189 |  |  | 435509 | 1656 |
|  | 1982 | 163120 | 363 |  | - |  |  |  |  |  |  | 174229 | 985 |  |  | 337349 | 1348 |
|  | 1983 | 278061 | 593 |  | - | - |  |  |  |  |  | 171361 | 957 |  |  | 449422 | 1550 |
|  | 1984 | 294365 | 628 |  |  |  |  |  |  |  |  | 176716 | 995 |  |  | 471081 | 1623 |
|  | 1985 | 299037 | 638 |  | - | - | - |  |  |  |  | 162403 | 923 |  |  | 461440 | 1561 |
|  | 1986 | 264849 | 556 |  | - | - |  |  |  |  |  | 191524 | 1042 |  |  | 456373 | 1598 |
|  | 1987 | 235703 | 491 |  | - | - |  |  |  |  |  | 153554 | 894 |  |  | 389257 | 1385 |
|  | 1988 | 217617 | 420 |  | - | - |  |  |  |  |  | 120367 | 656 |  |  | 337984 | 1076 |
|  | 1989 | 220170 | 436 |  | - | - | - |  |  |  |  | 80880 | 469 |  |  | 301050 | 905 |
|  | 1990 | 192500 | 385 |  | - | - |  |  |  |  |  | 91437 | 545 |  |  | 283937 | 930 |
|  | 1991 | 171041 | 342 |  | - | - | - |  |  |  |  | 92214 | 535 |  |  | 263255 | 877 |
|  | 1992 | 151291 | 301 |  | - |  | - |  |  |  |  | 92717 | 566 |  |  | 244008 | 867 |
|  | 1993 | 153407 | 312 | 62403 | 284 | 35147 | 327 |  |  |  |  |  |  |  |  | 250957 | 923 |
|  | 1994 |  | 415 |  | 319 |  | 262 |  |  |  |  |  |  |  |  |  | 996 |
|  | 1995 | 134341 | 249 | 71552 | 341 | 27104 | 249 |  |  |  |  |  | - |  |  | 232997 | 839 |
|  | 1996 | 110085 | 215 | 69389 | 322 | 27627 | 249 |  |  |  |  |  | - |  |  | 207101 | 786 |
|  | 1997 | 124387 | 241 | 52842 | 238 | 16448 | 151 |  |  |  |  |  | - |  |  | 193677 | 630 |
|  | 1998 | 162185 | 296 | 66767 | 306 | 15568 | 139 |  |  |  |  |  | - |  |  | 244520 | 741 |
|  | 1999 | 164905 | 318 | 70825 | 326 | 18669 | 167 |  |  |  |  |  | - |  |  | 254399 | 811 |
|  | 2000 | 250468 | 504 | 99934 | 454 | 24319 | 219 |  |  |  |  | - | - |  |  | 374721 | 1177 |
|  | 2001 | 207934 | 417 | 117759 | 554 | 33047 | 295 |  |  |  |  |  | - |  |  | 358740 | 1266 |
|  | 2002 | 127039 | 249 | 98055 | 471 | 33013 | 299 |  |  |  |  | - | - |  |  | 258107 | 1019 |
|  | 2003 | 185574 | 363 | 87993 | 410 | 31099 | 298 |  |  |  |  | - | - |  |  | 304666 | 1071 |
|  | 2004 | 108645 | 207 | 77343 | 371 | 23173 | 206 |  |  |  |  |  | - |  |  | 209161 | 784 |
|  | 2005 | 165900 | 307 | 69488 | 320 | 27507 | 261 |  |  |  |  | - | - |  |  | 262895 | 888 |
|  | 2006 | 142218 | 261 | 99401 | 453 | 23529 | 218 |  |  |  |  |  | - |  |  | 265148 | 932 |
|  | 2007 | 78165 | 140 | 79146 | 363 | 28896 | 264 |  |  |  |  |  | - |  |  | 186207 | 767 |
|  | 2008 | 89228 | 170 | 69027 | 314 | 34124 | 322 |  |  |  |  | - | - |  |  | 192379 | 807 |
|  | 2009 | 73045 | 135 | 53725 | 241 | 23663 | 219 |  |  |  |  | - | - |  |  | 150433 | 595 |
|  | 2010 | 98490 | 184 | 56260 | 250 | 22310 | 208 |  |  |  |  | - | - |  |  | 177060 | 642 |
|  | 2011 | 71597 | 140 | 81351 | 374 | 20270 | 183 |  |  |  |  | - | - |  |  | 173218 | 696 |
|  | 2012 | 81638 | 162 | 63985 | 289 | 26689 | 245 |  |  |  |  |  | - |  |  | 172312 | 696 |

## Russia

| Russia | 1987 |  |  | 27135 |  |  |  |  |  |  |  |  |  | 2521 |  | 137011 | 564 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1988 | 53158 |  | 33395 |  | 10256 |  | 294 |  | 25 |  | - | - | 2937 |  | 100065 | 420 |
|  | 1989 | 78023 | - | 23123 |  | 4118 |  | 26 |  | 0 | - | - | - | 2187 | - | 107477 | 364 |
|  | 1990 | 70595 |  | 20633 |  | 2919 |  | 101 |  | 0 | - | - | - | 2010 |  | 96258 | 313 |
|  | 1991 | 40603 | - | 12458 |  | 3060 |  | 650 | - | 0 | - | - | - | 1375 | - | 58146 | 215 |
|  | 1992 | 34021 | - | 8880 |  | 3547 |  | 180 | - | 0 | - | - | - | 824 | - | 47452 | 167 |
|  | 1993 | 28100 | - | 11780 |  | 4280 | - | 377 | - | 0 | - | - | - | 1470 | - | 46007 | 139 |
|  | 1994 | 30877 | - | 10879 | - | 2183 | - | 51 | - | 0 | - | - | - | 555 | - | 44545 | 141 |
|  | 1995 | 27775 | 62 | 9642 | 50 | 1803 | 15 | 6 | 0 | 0 | 0 | - | - | 385 | 2 | 39611 | 129 |
|  | 1996 | 33878 | 79 | 7395 | 42 | 1084 | 9 | 40 | 1 | 0 | 0 | - | - | 41 | 1 | 42438 | 131 |
|  | 1997 | 31857 | 72 | 5837 | 28 | 672 | 6 | 38 | 1 | 0 | 0 | - | - | 559 | 3 | 38963 | 110 |
|  | 1998 | 34870 | 92 | 6815 | 33 | 181 | 2 | 28 | 0 | 0 | 0 | - | - | 638 | 3 | 42532 | 130 |
|  | 1999 | 24016 | 66 | 5317 | 25 | 499 | 5 | 0 | 0 | 0 | 0 | - | - | 1131 | 6 | 30963 | 102 |
|  | 2000 | 27702 | 75 | 7027 | 34 | 500 | 5 | 3 | 0 | 0 | 0 | - | - | 1853 | 9 | 37085 | 123 |
|  | 2001 | 26472 | 61 | 7505 | 39 | 1036 | 10 | 30 | 0 | 0 | 0 | - | - | 922 | 5 | 35965 | 115 |
|  | 2002 | 24588 | 60 | 8720 | 43 | 1284 | 12 | 3 | 0 | 0 | 0 | - | - | 480 | 3 | 35075 | 118 |
|  | 2003 | 22014 | 50 | 8905 | 42 | 1206 | 12 | 20 | 0 | 0 | 0 | - | - | 634 | 4 | 32779 | 107 |
|  | 2004 | 17105 | 39 | 6786 | 33 | 880 | 7 | 0 | 0 | 0 | 0 | - | - | 529 | 3 | 25300 | 82 |
|  | 2005 | 16591 | 39 | 7179 | 33 | 989 | 8 | 1 | 0 | 0 | 0 | - | - | 439 | 3 | 25199 | 82 |
|  | 2006 | 22412 | 54 | 5392 | 28 | 759 | 6 | 0 | 0 | 0 | 0 | - | - | 449 | 3 | 29012 | 91 |
|  | 2007 | 12474 | 30 | 4377 | 23 | 929 | 7 | 0 | 0 | 0 | 0 | - | - | 277 | 2 | 18057 | 62 |
|  | 2008 | 13404 | 28 | 8674 | 39 | 669 | 4 | 8 | 0 | 0 | 0 | - | - | 312 | 2 | 23067 | 73 |
|  | 2009 | 13580 | 30 | 7215 | 35 | 720 | 5 | 36 | 0 | 0 | 0 | - | - | 173 | 1 | 21724 | 71 |
|  | 2010 | 14834 | 33 | 9821 | 48 | 844 | 6 | 49 | 0 | 0 | 0 | - | - | 186 | 1 | 25734 | 88 |
|  | 2011 | 13779 | 31 | 9030 | 44 | 747 | 5 | 51 | 0 | 0 | 0 | - | - | 171 | 1 | 23778 | 82 |
|  | 2012 | 17484 | 42 | 6560 | 34 | 738 | 5 | 53 | 0 | 0 | 0 | - | $-$ | 173 | 1 | 25008 | 83 |

Ireland

| Country | Year | 1SW |  | 2SW |  | 3SW |  | 4SW |  | 5SW |  | MSW (1) |  | PS |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt |
| Ireland | 1980 | 248333 | 745 |  |  |  |  |  |  |  |  | 39608 | 202 |  |  | 287941 | 947 |
|  | 1981 | 173667 | 521 |  |  |  |  |  |  |  |  | 32159 | 164 |  |  | 205826 | 685 |
|  | 1982 | 310000 | 930 |  |  |  |  |  |  |  |  | 12353 | 63 |  |  | 322353 | 993 |
|  | 1983 | 502000 | 1506 |  |  |  |  |  |  |  |  | 29411 | 150 |  |  | 531411 | 1656 |
|  | 1984 | 242666 | 728 |  |  |  |  |  |  |  |  | 19804 | 101 |  |  | 262470 | 829 |
|  | 1985 | 498333 | 1495 |  |  |  |  |  |  |  |  | 19608 | 100 |  |  | 517941 | 1595 |
|  | 1986 | 498125 | 1594 |  |  |  |  |  |  |  |  | 28335 | 136 |  |  | 526460 | 1730 |
|  | 1987 | 358842 | 1112 |  |  |  |  |  |  |  |  | 27609 | 127 |  |  | 386451 | 1239 |
|  | 1988 | 559297 | 1733 |  |  |  |  |  |  |  |  | 30599 | 141 |  |  | 589896 | 1874 |
|  | 1989 |  |  |  |  |  |  |  |  |  |  |  | - |  |  | 330558 | 1079 |
|  | 1990 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 188890 | 567 |
|  | 1991 | - |  |  |  |  |  |  |  |  |  |  | - |  |  | 135474 | 404 |
|  | 1992 | - |  |  |  |  |  |  |  |  |  |  | - |  |  | 235435 | 631 |
|  | 1993 | - |  |  |  |  |  |  |  |  |  | - | - |  |  | 200120 | 541 |
|  | 1994 | - |  |  |  |  |  |  |  |  |  |  | - |  |  | 286266 | 804 |
|  | 1995 | - |  |  |  |  |  |  |  |  |  | - | - |  |  | 288225 | 790 |
|  | 1996 | - |  |  |  |  |  |  |  |  |  |  | - |  |  | 249623 | 685 |
|  | 1997 | - | - |  |  |  |  |  |  |  |  | - | - |  |  | 209214 | 570 |
|  | 1998 | - |  |  |  |  |  |  |  |  |  |  | - |  |  | 237663 | 624 |
|  | 1999 | - |  |  |  |  |  |  |  |  |  | - | - |  |  | 180477 | 515 |
|  | 2000 | - |  |  |  |  |  |  |  |  |  | - | - |  |  | 228220 | 621 |
|  | 2001 | - |  |  |  |  |  |  |  |  |  | - | - |  |  | 270963 | 730 |
|  | 2002 | - |  |  |  |  |  |  |  |  |  | - | - |  |  | 256808 | 682 |
|  | 2003 | - | - |  |  |  |  |  |  |  |  | - | - |  |  | 204145 | 551 |
|  | 2004 | - |  |  |  |  |  |  |  |  |  |  | - |  |  | 180953 | 489 |
|  | 2005 | - | - |  |  |  |  |  |  |  |  | - | - |  |  | 156308 | 422 |
|  | 2006 | - |  |  |  |  |  |  |  |  |  |  | - |  |  | 120834 | 326 |
|  | 2007 | - |  |  |  |  |  |  |  |  |  | - | - |  |  | 30946 | 84 |
|  | 2008 | - |  |  |  |  |  |  |  |  |  |  | - |  |  | 33200 | 89 |
|  | 2009 | - | - |  |  |  |  |  |  |  |  | - | - |  |  | 25170 | 68 |
|  | 2010 | - | - |  |  |  |  |  |  |  |  | - | - |  |  | 36508 | 99 |
|  | 2011 | - | - |  |  |  |  |  |  |  |  | - | - |  |  | 32308 | 87 |
|  | 2012 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 32599 | 88 |

## UK(England and Wales)

| UK | 1985 | 62815 |  |  |  |  |  |  |  |  |  | 32716 |  |  |  | 95531 | 361 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (England \& Wales) | 1986 | 68759 | - | - | - | - | - | - | - | - | - | 42035 | - | - | - | 110794 | 430 |
| Wales | 1987 | 56739 | - | - | - | - | - | - | - | - | - | 26700 | - | - |  | 83439 | 302 |
|  | 1988 | 76012 | - | - | - | - | - | - | - | - | - | 34151 | - | - | - | 110163 | 395 |
|  | 1989 | 54384 | - | - | - | - | - | - | - | - | - | 29284 | - | - | - | 83668 | 296 |
|  | 1990 | 45072 | - | - | - | - | - | - | - | - | - | 41604 | - | - |  | 86676 | 338 |
|  | 1991 | 36671 | - | - | - | - | - | - | - | - | - | 14978 | - | - | - | 51649 | 200 |
|  | 1992 | 34331 | - | - | - | - | - | - | - | - | - | 10255 | - | - | - | 44586 | 171 |
|  | 1993 | 56033 | - | - | - | - | - | - | - | - | - | 13144 | - | - | - | 69177 | 248 |
|  | 1994 | 67853 | - | - | - | - | - | - | - | - | - | 20268 | - | - | - | 88121 | 324 |
|  | 1995 | 57944 | - | - | - | - | - | - | - | - | - | 22534 | - | - | - | 80478 | 295 |
|  | 1996 | 30352 | - | - | - | - | - | - | - | - | - | 16344 | - | - | - | 46696 | 183 |
|  | 1997 | 30203 | - | - | - | - | - | - | - | - | - | 11171 | - | - | - | 41374 | 142 |
|  | 1998 | 30641 | - | - | - | - | - | - | - | - | - | 6276 | - | - | - | 36917 | 123 |
|  | 1999 | 27944 | - | - | - | - | - | - | - | - | - | 13150 | - | - | - | 41094 | 150 |
|  | 2000 | 48153 | - | - | - | - | - | - | - | - | - | 12800 | - | - | - | 60953 | 219 |
|  | 2001 | 38993 | - | - | - | - | - | - | - | - | - | 12314 | - | - | - | 51307 | 184 |
|  | 2002 | 34708 | - | - | - | - | - | - | - | - | - | 10961 | - | - | - | 45669 | 161 |
|  | 2003 | 14878 | - | - | - | - | - | - | - | - | - | 7328 | - | - | - | 22206 | 89 |
|  | 2004 | 24753 | - | - | - | - | - | - | - | - | - | 5806 | - | - | - | 30559 | 111 |
|  | 2005 | 19622 | - | - | - | - | - | - | - | - | - | 6541 | - | - | - | 26162 | 97 |
|  | 2006 | 16983 | - | - | - | - | - | - | - | - | - | 5073 | - | - | - | 22056 | 80 |
|  | 2007 | 15540 | - | - | - | - | - | - | - | - | - | 4383 | - | - | - | 19923 | 67 |
|  | 2008 | 14277 | - | - | - | - | - | - | - | - | - | 4759 | - | - | - | 19036 | 64 |
|  | 2009 | 10015 | - | - | - | - | - | - | - | - | - | 3895 | - | - |  | 13910 | 54 |
|  | 2010 | 25502 | - | - | - | - | - | - | - | - | - | 7193 | - | - | - | 32695 | 109 |
|  | 2011 | 19708 | - | - | - | - | - | - | - | - | - | 14867 | - | - | - | 34575 | 136 |
|  | 2012 | 7496 |  |  |  |  |  |  |  |  |  | 7203 |  |  |  | 14699 | 57 |

## UK(Scotland)

| Country | Year | 1SW |  | 2SW |  | 3SW |  | 4SW |  | 5SW |  | MSW (1) |  | PS |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt | No. | Wt |
| UK (Scotland) | 1982 | 208061 | 496 |  |  |  |  |  |  |  |  | 128242 | 596 |  |  | 336303 | 1092 |
|  | 1983 | 209617 | 549 |  |  |  |  |  |  |  |  | 145961 | 672 |  |  | 355578 | 1221 |
|  | 1984 | 213079 | 509 |  |  |  |  |  |  |  |  | 107213 | 504 |  |  | 320292 | 1013 |
|  | 1985 | 158012 | 399 |  |  |  |  |  |  |  |  | 114648 | 514 |  |  | 272660 | 913 |
|  | 1986 | 202838 | 525 |  |  |  |  |  |  |  |  | 148197 | 744 |  |  | 351035 | 1269 |
|  | 1987 | 164785 | 419 |  |  |  |  |  |  |  |  | 103994 | 503 |  |  | 268779 | 922 |
|  | 1988 | 149098 | 381 |  |  |  |  |  |  |  |  | 112162 | 501 |  |  | 261260 | 882 |
|  | 1989 | 174941 | 431 |  |  |  |  |  |  |  |  | 103886 | 464 |  |  | 278827 | 895 |
|  | 1990 | 81094 | 201 |  |  |  |  |  |  |  |  | 87924 | 423 |  |  | 169018 | 624 |
|  | 1991 | 73608 | 177 |  |  |  |  |  |  |  |  | 65193 | 285 |  |  | 138801 | 462 |
|  | 1992 | 101676 | 238 |  |  |  |  |  |  |  |  | 82841 | 361 |  |  | 184517 | 600 |
|  | 1993 | 94517 | 227 |  |  |  |  |  |  |  |  | 71726 | 320 |  |  | 166243 | 547 |
|  | 1994 | 99479 | 248 |  |  |  |  |  |  |  |  | 85404 | 400 |  |  | 184883 | 648 |
|  | 1995 | 89971 | 224 |  |  |  |  |  |  |  |  | 78511 | 364 |  |  | 168482 | 588 |
|  | 1996 | 66465 | 160 |  |  |  |  |  |  |  |  | 57998 | 267 |  |  | 124463 | 427 |
|  | 1997 | 46866 | 114 |  |  |  |  |  |  |  |  | 40459 | 182 |  |  | 87325 | 296 |
|  | 1998 | 53503 | 121 |  |  |  |  |  |  |  |  | 39264 | 162 |  |  | 92767 | 283 |
|  | 1999 | 25255 | 57 |  |  |  |  |  |  |  |  | 30694 | 143 |  |  | 55949 | 199 |
|  | 2000 | 44033 | 114 |  |  |  |  |  |  |  |  | 36767 | 161 |  |  | 80800 | 275 |
|  | 2001 | 42586 | 101 |  |  |  |  |  |  |  |  | 34926 | 150 |  |  | 77512 | 251 |
|  | 2002 | 31385 | 73 |  |  |  |  |  |  |  |  | 26403 | 118 |  |  | 57788 | 191 |
|  | 2003 | 29598 | 71 |  |  |  |  |  |  |  |  | 27588 | 122 |  |  | 57091 | 192 |
|  | 2004 | 37631 | 88 |  |  |  |  |  |  |  |  | 36856 | 159 |  |  | 74033 | 245 |
|  | 2005 | 39093 | 91 |  |  |  |  |  |  |  |  | 28666 | 126 |  |  | 67117 | 215 |
|  | 2006 | 36668 | 75 |  |  |  |  |  |  |  |  | 27620 | 118 |  |  | 63848 | 192 |
|  | 2007 | 32335 | 71 |  |  |  |  |  |  |  |  | 24098 | 100 |  |  | 56433 | 171 |
|  | 2008 | 23431 | 51 |  |  |  |  |  |  |  |  | 25745 | 110 |  |  | 49176 | 161 |
|  | 2009 | 18189 | 37 |  |  |  |  |  |  |  |  | 19185 | 83 |  |  | 37374 | 121 |
|  | 2010 | 33426 | 69 |  |  |  |  |  |  |  |  | 26988 | 111 |  |  | 60414 | 180 |
|  | 2011 | 15706 | 33 |  |  |  |  |  |  |  |  | 28496 | 126 |  |  | 44202 | 159 |
|  | 2012 | 20429 | 42 |  |  |  |  |  |  |  |  | 20824 | 88 |  |  | 41253 | 130 |

## France



Spain


1. MSW includes all sea ages $>1$, when this cannot be broken down. Different methods are used to separate 1 SW and MSW salmon in different countries:

Scale reading: Faroe Islands, Finland (1996 onwards), France, Russia, USA and West Greenland.
Size (split weight/length): Canada ( 2.7 kg for nets; 63 cm for rods), Finland up until 1995 ( 3 kg ).
Iceland (various splits used at different times and places), Norway ( 3 kg ), UK Scotland ( 3 kg in some places and 3.7 kg in others). All countries except Scotland report no problems with using weight to categorise catches into sea age classes; mis-classification may be very high in some years. In Norway, catches shown as 3 SW refer to salmon of $3 S W$ or greater.
2. Based on catches in Asturias (80-90\% of total catch). No data for 2008, previous year data are used.

## Annex 5: Input data for NEAC Pre Fishery Abundance analysis using Monte Carlo simulation

Finland

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
|  |  |  |  |  |  |  |  |  |  |  |
| 1971 | 8,422 | 8,538 | 35 | 5 | 35 | 5 | 50 | 10 | 55 | 15 |
| 1972 | 13,160 | 13,341 | 35 | 5 | 35 | 5 | 50 | 10 | 55 | 15 |
| 1973 | 11,969 | 15,958 | 35 | 5 | 35 | 5 | 50 | 10 | 55 | 15 |
| 1974 | 23,709 | 23,709 | 35 | 5 | 35 | 5 | 50 | 10 | 55 | 15 |
| 1975 | 16,527 | 26,417 | 35 | 5 | 35 | 5 | 50 | 10 | 55 | 15 |
| 1976 | 11,323 | 21,719 | 35 | 5 | 35 | 5 | 50 | 10 | 55 | 15 |
| 1977 | 5,807 | 13,227 | 35 | 5 | 35 | 5 | 50 | 10 | 55 | 15 |
| 1978 | 7,902 | 8,452 | 35 | 5 | 35 | 5 | 50 | 10 | 55 | 15 |
| 1979 | 9,249 | 7,390 | 35 | 5 | 35 | 5 | 50 | 10 | 45 | 15 |
| 1980 | 4,792 | 8,938 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1981 | 7,386 | 9,835 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1982 | 2,163 | 12,826 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1983 | 10,680 | 13,990 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1984 | 11,942 | 13,262 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1985 | 18,039 | 10,339 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1986 | 16,389 | 9,028 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1987 | 20,950 | 11,290 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1988 | 10,019 | 7,231 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1989 | 28,091 | 10,011 | 25 | 5 | 25 | 5 | 60 | 10 | 55 | 15 |
| 1990 | 26,646 | 12,562 | 25 | 5 | 25 | 5 | 60 | 10 | 55 | 15 |
| 1991 | 32,423 | 15,136 | 25 | 5 | 25 | 5 | 60 | 10 | 55 | 15 |
| 1992 | 42,965 | 16,158 | 25 | 5 | 25 | 5 | 60 | 10 | 55 | 15 |
| 1993 | 30,197 | 18,720 | 25 | 5 | 25 | 5 | 60 | 10 | 55 | 15 |
| 1994 | 12,016 | 15,521 | 25 | 5 | 25 | 5 | 60 | 10 | 55 | 15 |
| 1995 | 11,801 | 9,634 | 25 | 5 | 25 | 5 | 60 | 10 | 55 | 15 |
| 1996 | 22,799 | 6,956 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1997 | 19,481 | 10,083 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1998 | 22,460 | 8,497 | 25 | 5 | 25 | 5 | 50 | 10 | 45 | 15 |
| 1999 | 38,687 | 8,854 | 25 | 5 | 25 | 5 | 60 | 10 | 50 | 10 |
| 2000 | 40,654 | 19,707 | 25 | 5 | 25 | 5 | 60 | 10 | 50 | 10 |
| 2001 | 18,372 | 28,337 | 25 | 5 | 25 | 5 | 60 | 10 | 50 | 10 |
| 2002 | 10,757 | 22,717 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| 2003 | 12,699 | 16,093 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| 2004 | 4,912 | 7,718 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| 2005 | 12,499 | 5,969 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| 2006 | 23,727 | 10,473 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| 2007 | 4,407 | 14,878 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| 2008 | 4,539 | 14,165 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| 2009 | 9,260 | 6,600 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| 2010 | 8627 | 10434 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| 2011 | 10554 | 8204 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| 2012 | 22902 | 10649 | 25 | 5 | 25 | 5 | 50 | 10 | 50 | 10 |
| $\begin{array}{r} \mathrm{M}(\min )= \\ \mathrm{M}(\max )= \end{array}$ | 0.020 |  | Return time (m)= |  | 1SW(min) | 7 9 | MSW (min) MSW (max) | 16 18 |  |  |

France

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
|  |  |  |  |  |  |  |  |  |  |  |
| 1971 | 1,740 | 4,060 |  |  |  |  | 3.5 | 1.5 | 37.5 | 12.5 |
| 1972 | 3,480 | 8,120 |  |  |  |  | 3.5 | 1.5 | 37.5 | 12.5 |
| 1973 | 2,130 | 4,970 |  |  |  |  | 3.5 | 1.5 | 37.5 | 12.5 |
| 1974 | 990 | 2,310 |  |  |  |  | 3.5 | 1.5 | 37.5 | 12.5 |
| 1975 | 1,980 | 4,620 |  |  |  |  | 3.5 | 1.5 | 37.5 | 12.5 |
| 1976 | 1,820 | 3,380 |  |  |  |  | 3.5 | 1.5 | 37.5 | 12.5 |
| 1977 | 1,400 | 2,600 |  |  |  |  | 3.5 | 1.5 | 37.5 | 12.5 |
| 1978 | 1,435 | 2,665 |  |  |  |  | 3.5 | 1.5 | 37.5 | 12.5 |
| 1979 | 1,645 | 3,055 |  |  |  |  | 3.5 | 1.5 | 37.5 | 12.5 |
| 1980 | 3,430 | 6,370 |  |  |  |  | 3.5 | 1.5 | 37.5 | 12.5 |
| 1981 | 2,720 | 4,080 |  |  |  |  | 3.5 | 1.5 | 35 | 15 |
| 1982 | 1,680 | 2,520 |  |  |  |  | 3.5 | 1.5 | 35 | 15 |
| 1983 | 1,800 | 2,700 |  |  |  |  | 3.5 | 1.5 | 35 | 15 |
| 1984 | 2,960 | 4,440 |  |  |  |  | 3.5 | 1.5 | 35 | 15 |
| 1985 | 1,100 | 3,330 |  |  |  |  | 3.5 | 1.5 | 35 | 15 |
| 1986 | 3,400 | 3,400 |  |  |  |  | 7 | 5 | 35 | 15 |
| 1987 | 6,013 | 1,806 |  |  |  |  | 7 | 5 | 35 | 15 |
| 1988 | 2,063 | 4,964 |  |  |  |  | 7 | 5 | 35 | 15 |
| 1989 | 1,124 | 2,282 |  |  |  |  | 7 | 5 | 35 | 15 |
| 1990 | 1,886 | 2,332 |  |  |  |  | 7 | 5 | 35 | 15 |
| 1991 | 1,362 | 2,125 |  |  |  |  | 7 | 5 | 35 | 15 |
| 1992 | 2,490 | 2,671 |  |  |  |  | 7 | 5 | 35 | 15 |
| 1993 | 3,581 | 1,254 |  |  |  |  | 7 | 5 | 35 | 15 |
| 1994 | 2,810 | 2,290 |  |  |  |  | 7 | 5 | 30 | 10 |
| 1995 | 1,669 | 1,095 |  |  |  |  | 12.5 | 7.5 | 30 | 10 |
| 1996 | 2,063 | 1,943 |  |  |  |  | 12.5 | 7.5 | 30 | 10 |
| 1997 | 1,060 | 1,001 |  |  |  |  | 12.5 | 7.5 | 30 | 10 |
| 1998 | 2,065 | 846 |  |  |  |  | 12.5 | 7.5 | 30 | 10 |
| 1999 | 690 | 1,831 |  |  |  |  | 12.5 | 7.5 | 30 | 10 |
| 2000 | 1,792 | 1,277 |  |  |  |  | 12.5 | 7.5 | 30 | 10 |
| 2001 | 1,544 | 1,489 |  |  |  |  | 12.5 | 7.5 | 30 | 10 |
| 2002 | 2,423 | 1,065 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| 2003 | 1,598 | 1,540 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| 2004 | 1,927 | 2,880 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| 2005 | 1,256 | 1,771 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| 2006 | 1,763 | 1,785 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| 2007 | 1,378 | 1,685 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| 2008 | 1,365 | 1,865 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| 2009 | 487 | 975 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| 2010 | 1658 | 821 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| 2011 | 1162 | 2142 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| 2012 | 1010 | 1669 | 30 | 10 | 23 | 8 | 13 | 8 | 30 | 10 |
| $\begin{array}{r} \mathrm{M}(\min )= \\ \mathrm{M}(\max )= \end{array}$ | 0.020 0.040 |  |  | me (m) $=$ | 1SW(min) 1SW(max) | 7 9 | MSW (min) MSW (max) | 16 18 |  |  |

Iceland (South and West)

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
| 1971 | 30,618 | 16,749 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1972 | 24,832 | 25,733 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1973 | 26,624 | 23,183 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1974 | 18,975 | 20,017 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1975 | 29,428 | 21,266 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1976 | 23,233 | 18,379 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1977 | 23,802 | 17,919 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1978 | 31,199 | 23,182 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1979 | 28,790 | 14,840 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1980 | 13,073 | 20,855 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1981 | 16,890 | 13,919 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1982 | 17,331 | 9,826 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1983 | 21,923 | 16,423 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1984 | 13,476 | 13,923 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1985 | 21,822 | 10,097 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1986 | 35,891 | 8,423 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1987 | 22,302 | 7,480 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1988 | 40,028 | 8,523 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1989 | 22,377 | 7,607 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1990 | 20,584 | 7,548 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1991 | 22,711 | 7,519 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1992 | 26,006 | 8,479 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1993 | 25,479 | 4,155 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1994 | 20,985 | 6,736 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1995 | 25,371 | 6,777 | 12.5 | 2.5 | 12.5 | 2.5 | 50 | 10 | 70 | 10 |
| 1996 | 21,913 | 4,364 | 12.5 | 2.5 | 12.5 | 2.5 | 50 | 10 | 70 | 10 |
| 1997 | 16,007 | 4,910 | 12.5 | 2.5 | 12.5 | 2.5 | 50 | 10 | 70 | 10 |
| 1998 | 21,900 | 3,037 | 12.5 | 2.5 | 12.5 | 2.5 | 50 | 10 | 70 | 10 |
| 1999 | 17,448 | 5,757 | 12.5 | 2.5 | 12.5 | 2.5 | 49 | 10 | 68 | 10 |
| 2000 | 15,502 | 1,519 | 12.5 | 2.5 | 12.5 | 2.5 | 49 | 10 | 66 | 10 |
| 2001 | 13,586 | 2,707 | 12.5 | 2.5 | 12.5 | 2.5 | 48 | 10 | 67 | 10 |
| 2002 | 16,952 | 2,845 | 13 | 3 | 13 | 3 | 48 | 10 | 65 | 10 |
| 2003 | 20,271 | 4,751 | 13 | 3 | 13 | 3 | 48 | 10 | 68 | 10 |
| 2004 | 20,319 | 3,784 | 13 | 3 | 13 | 3 | 48 | 10 | 67 | 10 |
| 2005 | 29,969 | 3,241 | 13 | 3 | 13 | 3 | 48 | 10 | 65 | 10 |
| 2006 | 21,153 | 2,689 | 13 | 3 | 13 | 3 | 48 | 10 | 65 | 10 |
| 2007 | 23,728 | 1,679 | 13 | 3 | 13 | 3 | 47 | 9 | 66 | 10 |
| 2008 | 28,774 | 1,659 | 13 | 3 | 13 | 3 | 47 | 10 | 57 | 10 |
| 2009 | 33,190 | 2,838 | 13 | 3 | 13 | 3 | 48 | 10 | 63 | 10 |
| 2010 | 33,318 | 6,061 | 13 | 3 | 13 | 3 | 47 | 10 | 65 | 10 |
| 2011 | 23436 | 2934 | 13 | 3 | 13 | 3 | 47 | 10 | 62 | 10 |
| 2012 | 14975 | 2398 | 13 | 3 | 13 | 3 | 48 | 10 | 59 | 10 |
| $\begin{array}{r} \mathrm{M}(\min )= \\ \mathrm{M}(\max )= \end{array}$ | 0.020 0.040 |  |  | me (m) $=$ | 1SW(min) 1SW(max) | 7 9 | MSW (min) MSW (max) | 16 18 |  |  |

Iceland (North and East)

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
|  |  |  |  |  |  |  |  |  |  |  |
| 1971 | 4,610 | 6,625 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1972 | 4,223 | 10,337 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1973 | 5,060 | 9,672 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1974 | 5,047 | 9,176 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1975 | 6,152 | 10,136 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1976 | 6,184 | 8,350 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1977 | 8,597 | 11,631 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1978 | 8,739 | 14,998 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1979 | 8,363 | 9,897 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1980 | 1,268 | 13,784 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1981 | 6,528 | 4,827 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1982 | 3,007 | 5,539 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1983 | 4,437 | 4,224 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1984 | 1,611 | 5,447 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1985 | 11,116 | 3,511 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1986 | 13,827 | 9,569 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1987 | 8,145 | 9,908 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1988 | 11,775 | 6,381 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1989 | 6,342 | 5,414 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1990 | 4,752 | 5,709 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1991 | 6,900 | 3,965 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1992 | 12,996 | 5,903 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1993 | 10,689 | 6,672 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1994 | 3,414 | 5,656 | 2 | 1 | 2 | 1 | 50 | 10 | 70 | 10 |
| 1995 | 8,776 | 3,511 | 12.5 | 2.5 | 12.5 | 2.5 | 50 | 10 | 70 | 10 |
| 1996 | 4,681 | 4,605 | 12.5 | 2.5 | 12.5 | 2.5 | 50 | 10 | 70 | 10 |
| 1997 | 6,406 | 2,594 | 12.5 | 2.5 | 12.5 | 2.5 | 50 | 10 | 70 | 10 |
| 1998 | 10,905 | 3,780 | 12.5 | 2.5 | 12.5 | 2.5 | 50 | 10 | 70 | 10 |
| 1999 | 5,326 | 4,030 | 12.5 | 2.5 | 12.5 | 2.5 | 48 | 10 | 65 | 10 |
| 2000 | 5,595 | 2,324 | 12.5 | 2.5 | 12.5 | 2.5 | 48 | 10 | 64 | 10 |
| 2001 | 4,976 | 2,587 | 12.5 | 2.5 | 12.5 | 2.5 | 47 | 10 | 62 | 10 |
| 2002 | 8,437 | 2,366 | 13 | 3 | 13 | 3 | 46 | 10 | 60 | 10 |
| 2003 | 4,478 | 2,194 | 13 | 3 | 13 | 3 | 46 | 10 | 53 | 10 |
| 2004 | 11,823 | 2,239 | 13 | 3 | 13 | 3 | 45 | 10 | 55 | 10 |
| 2005 | 10,297 | 2,726 | 13 | 3 | 13 | 3 | 44 | 10 | 54 | 10 |
| 2006 | 11,082 | 2,179 | 13 | 3 | 13 | 3 | 45 | 10 | 45 | 10 |
| 2007 | 8,046 | 1,672 | 13 | 3 | 13 | 3 | 44 | 10 | 36 | 10 |
| 2008 | 7,021 | 2,693 | 13 | 3 | 13 | 3 | 42 | 10 | 45 | 10 |
| 2009 | 10,779 | 1,735 | 13 | 3 | 13 | 3 | 40 | 10 | 36 | 10 |
| 2010 | 8,621 | 2,602 | 13 | 3 | 13 | 3 | 40 | 10 | 38 | 10 |
| 2011 | 6759 | 2596 | 13 | 3 | 13 | 3 | 38 | 10 | 34 | 10 |
| 2012 | 3058 | 1265 | 13 | 3 | 13 | 3 | 42 | 10 | 34 | 10 |
| $\begin{aligned} \mathrm{M}(\min ) & = \\ \mathrm{M}(\max ) & = \end{aligned}$ | 0.020 0.040 |  | Ret | me (m)= | 1SW(min) 1SW(max) | 7 9 | MSW (min) MSW(max) | 16 18 |  |  |


| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  | Net catch |  | Catch \& Release |  | Spawner numbers small rivers |  | Spawner numbers closed rivers |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1sw | Msw | est | error | est | error | est | error | est | error | 1sw | Msw | 1sw | Msw | 1sw | Msw | 1sw | Msw |
| 1971 | 409,965 | 46,594 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1972 | 437,089 | 49,863 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1973 | 476,131 | 54,008 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1974 | 542,124 | 60,976 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1975 | 598,524 | 68,260 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1976 | 407,018 | 47,358 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1977 | 351,745 | 41,256 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1978 | 307,569 | 35,708 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1979 | 282,700 | 32,144 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1980 | 215,116 | 35,447 | 37.5 | 7.5 | 37.5 | 7.5 | 62.5 | 12.5 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1981 | 137,366 | 26,101 | 37.5 | 7.5 | 37.5 | 7.5 | 75.7 | 11.4 | 47.5 | 12.5 |  |  |  |  |  |  |  |  |
| 1982 | 269,847 | 11,754 | 37.5 | 7.5 | 37.5 | 7.5 | 71.9 | 10.8 | 36.7 | 8.3 |  |  |  |  |  |  |  |  |
| 1983 | 437,751 | 26,479 | 37.5 | 7.5 | 37.5 | 7.5 | 66.1 | 9.9 | 27.9 | 17.5 |  |  |  |  |  |  |  |  |
| 1984 | 224,872 | 20,685 | 37.5 | 7.5 | 37.5 | 7.5 | 64.6 | 9.7 | 43.5 | 6.5 |  |  |  |  |  |  |  |  |
| 1985 | 430,315 | 18,830 | 37.5 | 7.5 | 37.5 | 7.5 | 74.6 | 11.2 | 36.1 | 3.4 |  |  |  |  |  |  |  |  |
| 1986 | 443,701 | 27,111 | 37.5 | 7.5 | 37.5 | 7.5 | 68.7 | 10.3 | 46.0 | 9.0 |  |  |  |  |  |  |  |  |
| 1987 | 324,709 | 26,301 | 30 | 10 | 30 | 10 | 69.8 | 10.5 | 32.2 | 4.7 |  |  |  |  |  |  |  |  |
| 1988 | 391,475 | 22,067 | 30 | 10 | 30 | 10 | 62.0 | 9.3 | 37.4 | 5.6 |  |  |  |  |  |  |  |  |
| 1989 | 297,797 | 25,447 | 30 | 10 | 30 | 10 | 65.7 | 9.9 | 47.2 | 8.8 |  |  |  |  |  |  |  |  |
| 1990 | 172,098 | 15,549 | 30 | 10 | 30 | 10 | 60.7 | 9.1 | 59.9 | 6.1 |  |  |  |  |  |  |  |  |
| 1991 | 120,408 | 10,334 | 30 | 10 | 30 | 10 | 59.5 | 8.9 | 26.5 | 3.5 |  |  |  |  |  |  |  |  |
| 1992 | 182,255 | 15,456 | 30 | 10 | 30 | 10 | 62.1 | 9.3 | 51.5 | 3.8 |  |  |  |  |  |  |  |  |
| 1993 | 150,274 | 13,156 | 25 | 10 | 25 | 10 | 58.6 | 8.8 | 42.0 | 18.0 |  |  |  |  |  |  |  |  |
| 1994 | 234,126 | 20,506 | 25 | 10 | 25 | 10 | 71.4 | 10.7 | 40.5 | 2.5 |  |  |  |  |  |  |  |  |
| 1995 | 232,480 | 20,454 | 25 | 10 | 25 | 10 | 63.5 | 9.5 | 41.8 | 1.2 |  |  |  |  |  |  |  |  |
| 1996 | 203,920 | 18,021 | 25 | 10 | 25 | 10 | 59.9 | 9.0 | 55.1 | 3.2 |  |  |  |  |  |  |  |  |
| 1997 | 170,774 | 14,724 | 25 | 10 | 15 | 5 | 50.1 | 7.5 | 30.8 | 12.2 |  |  |  |  |  |  |  |  |
| 1998 | 191,868 | 17,269 | 25 | 10 | 15 | 5 | 53.7 | 8.1 | 61.9 | 1.4 |  |  |  |  |  |  |  |  |
| 1999 | 158,818 | 14,801 | 25 | 10 | 15 | 5 | 47.8 | 7.2 | 34.1 | 18.1 |  |  |  |  |  |  |  |  |
| 2000 | 199,827 | 16,848 | 25 | 10 | 15 | 5 | 43.2 | 6.5 | 31.0 | 4.5 |  |  |  |  |  |  |  |  |
| 2001 | 218,715 | 18,436 | 7.5 | 2.5 | 7.5 | 2.5 | 48 | 7.2 | 35 | 8 |  |  |  |  |  |  |  |  |
| 2002 | 198,719 | 16,702 | 8 | 3 | 8 | 3 | 50 | 7 | 28 | 8 |  |  |  |  |  |  |  |  |
| 2003 | 161,270 | 13,745 | 8 | 3 | 8 | 3 | 41 | 6 | 22 | 6 |  |  |  |  |  |  |  |  |
| 2004 | 142,251 | 12,299 | 8 | 3 | 8 | 3 | 50 | 8 | 35 | 8 |  |  |  |  |  |  |  |  |
| 2005 | 127,371 | 10,716 | 8 | 3 | 8 | 3 | 45 | 7 | 24 | 4 |  |  |  |  |  |  |  |  |
| 2006 | 101,938 | 9,740 |  | 3 | 8 | 3 | 47 | 7 | 30 | 14 |  |  |  |  |  |  |  |  |
| 2007 | 30,418 | 2,477 | 8 | 3 |  | 3 | 16 | 8 | 24 | 9 | 8334 | 679 | 12137 | 988 | 9548 | 777 | 40255 | 3278 |
| 2008 | 30,257 | 3,935 | 8 | 3 | 8 | 3 | 16 | 8 | 24 | 9 | 8253 | 650 | 10485 | 1492 | 12206 | 961 | 34382 | 4580 |
| 2009 | 24,184 | 4,756 | 8 | 3 | 8 | 3 | 16 | 8 | 24 | 9 | 6264 | 493 | 9799 | 1623 |  |  | 46570 | 4964 |
| 2010 | 33,211 | 3,297 |  | 3 | 8 | 3 | 16 | 8 | 24 | 9 | 13125 | 1034 | 13903 | 1255 |  |  | 35804 | 1504 |
| 2011 | 29117 | 3970 | 8 | 3 | 8 | 3 | 16 | 8 | 24 | 9 | 11071 | 902 | 11222 | 1530 |  |  | 33251 | 1208 |
| 2012 | 29979 | 4198 | 8 | 3 |  | 3 | 16 | 8 | 24 |  | 9542 | 777 | 10429 | 1463 |  |  | 32070 | 993 |
| $\mathrm{M}($ min $)=$ | 0.020 |  |  | ime (m)= |  |  | MSW(min) | 16 |  |  |  |  |  |  |  |  |  |  |
| $\mathrm{M}($ max $)=$ | 0.040 |  |  |  | 1SW(max) | 9 | MSW(max) | 18 |  |  |  |  |  |  |  |  |  |  |

Note: Net catch and spawner numbers from 2007 only (zero values for spawner numbers in small rivers from 2009 as these included in closed rivers)

## Norway (Southeast)

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| 1972 |  |  |  |  |  |  |  |  |  |  |
| 1973 |  |  |  |  |  |  |  |  |  |  |
| 1974 |  |  |  |  |  |  |  |  |  |  |
| 1975 |  |  |  |  |  |  |  |  |  |  |
| 1976 |  |  |  |  |  |  |  |  |  |  |
| 1977 |  |  |  |  |  |  |  |  |  |  |
| 1978 |  |  |  |  |  |  |  |  |  |  |
| 1979 |  |  |  |  |  |  |  |  |  |  |
| 1980 |  |  |  |  |  |  |  |  |  |  |
| 1981 |  |  |  |  |  |  |  |  |  |  |
| 1982 |  |  |  |  |  |  |  |  |  |  |
| 1983 | 9,039 | 9,004 | 50 | 10 | 50 | 10 | 70 | 10 | 65 | 10 |
| 1984 | 11,402 | 11,527 | 50 | 10 | 50 | 10 | 70 | 10 | 65 | 10 |
| 1985 | 18,699 | 11,883 | 50 | 10 | 50 | 10 | 70 | 10 | 65 | 10 |
| 1986 | 23,089 | 12,077 | 50 | 10 | 50 | 10 | 70 | 10 | 65 | 10 |
| 1987 | 19,601 | 14,179 | 50 | 10 | 50 | 10 | 70 | 10 | 65 | 10 |
| 1988 | 17,520 | 9,443 | 50 | 10 | 50 | 10 | 70 | 10 | 65 | 10 |
| 1989 | 23,965 | 12,254 | 50 | 10 | 50 | 10 | 65 | 10 | 60 | 10 |
| 1990 | 25,792 | 11,502 | 50 | 10 | 50 | 10 | 65 | 10 | 60 | 10 |
| 1991 | 21,064 | 10,753 | 50 | 10 | 50 | 10 | 65 | 10 | 60 | 10 |
| 1992 | 26,044 | 15,332 | 50 | 10 | 50 | 10 | 65 | 10 | 60 | 10 |
| 1993 | 23,070 | 12,596 | 40 | 10 | 40 | 10 | 65 | 10 | 60 | 10 |
| 1994 | 23,987 | 9,988 | 40 | 10 | 40 | 10 | 65 | 10 | 60 | 10 |
| 1995 | 21,847 | 11,630 | 40 | 10 | 40 | 10 | 65 | 10 | 60 | 10 |
| 1996 | 20,738 | 13,538 | 40 | 10 | 40 | 10 | 65 | 10 | 60 | 10 |
| 1997 | 21,121 | 7,756 | 35 | 10 | 35 | 10 | 60 | 10 | 60 | 10 |
| 1998 | 32,586 | 10,396 | 35 | 10 | 35 | 10 | 60 | 10 | 60 | 10 |
| 1999 | 23,904 | 6,664 | 35 | 10 | 35 | 10 | 60 | 10 | 60 | 10 |
| 2000 | 43,151 | 14,261 | 35 | 10 | 35 | 10 | 60 | 10 | 60 | 10 |
| 2001 | 47,339 | 19,210 | 35 | 10 | 35 | 10 | 60 | 10 | 60 | 10 |
| 2002 | 33,087 | 14,400 | 35 | 10 | 35 | 10 | 60 | 10 | 60 | 10 |
| 2003 | 33,371 | 20,648 | 30 | 10 | 30 | 10 | 60 | 10 | 60 | 10 |
| 2004 | 28,506 | 15,948 | 30 | 10 | 30 | 10 | 60 | 10 | 60 | 10 |
| 2005 | 40,628 | 14,628 | 30 | 10 | 30 | 10 | 60 | 10 | 60 | 10 |
| 2006 | 30,979 | 21,192 | 30 | 10 | 30 | 10 | 60 | 10 | 60 | 10 |
| 2007 | 15,735 | 18,130 | 30 | 10 | 30 | 10 | 60 | 10 | 60 | 10 |
| 2008 | 15,696 | 16,678 | 30 | 10 | 30 | 10 | 55 | 10 | 50 | 10 |
| 2009 | 15,584 | 11,995 | 30 | 10 | 30 | 10 | 55 | 10 | 50 | 10 |
| 2010 | 22,139 | 12,175 | 30 | 10 | 30 | 10 | 55 | 10 | 50 | 10 |
| 2011 | 15773 | 28589 | 30 | 10 | 30 | 10 | 50 | 10 | 40 | 10 |
| 2012 | 18582 | 23389 | 30 | 10 | 30 | 10 | 50 | 10 | 40 | 10 |
| $\begin{array}{r} \mathrm{M}(\min )= \\ \mathrm{M}(\max )= \end{array}$ | 0.020 0.040 |  |  | me $(\mathrm{m})=$ | 1SW(min) 1SW(max) | 7 9 | MSW (min) MSW (max) | 16 18 |  |  |

## Norway (Southwest)



Mid-Norway


| $M(\min )$ | $=0.020$ |
| ---: | :--- |
| $M(\max )$ | $=0.040$ |

Return time (m)=
1SW(min) 1SW(max)

[^1]16
18
MSW(max) $\quad 18$

## Norway North



Russia (Archangelsk and Karelia)

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
|  |  |  |  |  |  |  |  |  |  |  |
| 1971 | 134 | 16,592 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1972 | 116 | 14,434 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1973 | 169 | 20,924 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1974 | 170 | 21,137 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1975 | 140 | 17,398 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1976 | 111 | 13,781 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1977 | 78 | 9,722 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1978 | 82 | 10,134 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1979 | 112 | 13,903 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1980 | 156 | 19,397 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1981 | 68 | 8,394 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1982 | 71 | 8,797 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1983 | 48 | 11,938 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1984 | 21 | 10,680 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1985 | 454 | 11,183 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1986 | 12 | 12,291 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1987 | 647 | 8,734 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1988 | 224 | 9,978 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1989 | 989 | 10,245 | 10 | 5 | 10 | 5 | 60 | 20 | 60 | 20 |
| 1990 | 1,418 | 8,429 | 15 | 5 | 15 | 5 | 60 | 20 | 60 | 20 |
| 1991 | 421 | 8,725 | 20 | 5 | 20 | 5 | 60 | 20 | 60 | 20 |
| 1992 | 1,031 | 3,949 | 25 | 5 | 25 | 5 | 60 | 20 | 60 | 20 |
| 1993 | 196 | 4,251 | 30 | 5 | 30 | 5 | 60 | 20 | 60 | 20 |
| 1994 | 334 | 5,631 | 35 | 5 | 35 | 5 | 60 | 20 | 60 | 20 |
| 1995 | 386 | 5,214 | 45 | 5 | 45 | 5 | 60 | 20 | 60 | 20 |
| 1996 | 231 | 3,753 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 1997 | 721 | 3,351 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 1998 | 585 | 4,208 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 1999 | 299 | 3,101 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2000 | 514 | 3,382 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2001 | 363 | 2,348 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2002 | 1,676 | 2,439 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2003 | 893 | 2,041 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2004 | 990 | 3,761 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2005 | 1,349 | 4,915 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2006 | 2,183 | 2,841 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2007 | 1,618 | 2,621 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2008 | 332 | 2,496 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2009 | 252 | 2,214 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2010 | 397 | 3,823 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2011 | 313 | 2585 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| 2012 | 1332 | 2446 | 55 | 5 | 55 | 5 | 60 | 20 | 60 | 20 |
| $\begin{aligned} \mathrm{M}(\min ) & = \\ \mathrm{M}(\max ) & = \end{aligned}$ | 0.020 0.040 |  |  | me (m)= | 1SW(min) 1SW(max) | 7 8 | MSW (min) MSW(max) | 19 21 |  |  |

Russia (Kola Peninsula: Barents Sea Basin)

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
| 1971 | 4,892 | 5,979 | 15 | 5 | 15 | 5 | 45 | 5 | 45 | 5 |
| 1972 | 7,978 | 9,750 | 15 | 5 | 15 | 5 | 45 | 5 | 45 | 5 |
| 1973 | 9,376 | 11,460 | 15 | 5 | 15 | 5 | 40 | 5 | 40 | 5 |
| 1974 | 12,794 | 15,638 | 15 | 5 | 15 | 5 | 40 | 5 | 40 | 5 |
| 1975 | 13,872 | 13,872 | 15 | 5 | 15 | 5 | 45 | 5 | 45 | 5 |
| 1976 | 11,493 | 14,048 | 15 | 5 | 15 | 5 | 55 | 5 | 55 | 5 |
| 1977 | 7,257 | 8,253 | 15 | 5 | 15 | 5 | 50 | 5 | 50 | 5 |
| 1978 | 7,106 | 7,113 | 15 | 5 | 15 | 5 | 55 | 5 | 55 | 5 |
| 1979 | 6,707 | 3,141 | 15 | 5 | 15 | 5 | 40 | 5 | 40 | 5 |
| 1980 | 6,621 | 5,216 | 15 | 5 | 15 | 5 | 40 | 5 | 40 | 5 |
| 1981 | 4,547 | 5,973 | 15 | 5 | 15 | 5 | 40 | 5 | 40 | 5 |
| 1982 | 5,159 | 4,798 | 15 | 5 | 15 | 5 | 35 | 5 | 35 | 5 |
| 1983 | 8,504 | 9,943 | 15 | 5 | 15 | 5 | 35 | 5 | 35 | 5 |
| 1984 | 9,453 | 12,601 | 15 | 5 | 15 | 5 | 35 | 5 | 35 | 5 |
| 1985 | 6,774 | 7,877 | 15 | 5 | 15 | 5 | 35 | 5 | 35 | 5 |
| 1986 | 10,147 | 5,352 | 15 | 5 | 15 | 5 | 40 | 5 | 40 | 5 |
| 1987 | 8,560 | 5,149 | 15 | 5 | 15 | 5 | 40 | 5 | 40 | 5 |
| 1988 | 6,644 | 3,655 | 15 | 5 | 15 | 5 | 35 | 5 | 35 | 5 |
| 1989 | 13,424 | 6,787 | 15 | 5 | 15 | 5 | 40 | 5 | 40 | 5 |
| 1990 | 16,038 | 8,234 | 15 | 5 | 15 | 5 | 40 | 5 | 40 | 5 |
| 1991 | 4,550 | 7,568 | 15 | 5 | 15 | 5 | 30 | 5 | 30 | 5 |
| 1992 | 11,394 | 7,109 | 15 | 5 | 15 | 5 | 30 | 5 | 30 | 5 |
| 1993 | 8,642 | 5,690 | 15 | 5 | 15 | 5 | 30 | 5 | 30 | 5 |
| 1994 | 6,101 | 4,632 | 15 | 5 | 15 | 5 | 30 | 5 | 30 | 5 |
| 1995 | 6,318 | 3,693 | 15 | 5 | 15 | 5 | 30 | 5 | 30 | 5 |
| 1996 | 6,815 | 1,701 | 20 | 5 | 20 | 5 | 25 | 5 | 25 | 5 |
| 1997 | 3,564 | 867 | 25 | 5 | 25 | 5 | 15 | 5 | 15 | 5 |
| 1998 | 1,854 | 280 | 35 | 5 | 35 | 5 | 12.5 | 2.5 | 12.5 | 2.5 |
| 1999 | 1,510 | 424 | 40 | 5 | 40 | 5 | 7.5 | 2.5 | 7.5 | 2.5 |
| 2000 | 805 | 323 | 50 | 5 | 50 | 5 | 6 | 2 | 6 | 2 |
| 2001 | 591 | 241 | 60 | 5 | 60 | 5 | 3.5 | 1.5 | 3.5 | 1.5 |
| 2002 | 1,436 | 2,478 | 50 | 10 | 50 | 10 | 10 | 5 | 20 | 5 |
| 2003 | 1,938 | 1,095 | 50 | 10 | 50 | 10 | 10 | 5 | 20 | 5 |
| 2004 | 1,095 | 850 | 50 | 10 | 50 | 10 | 10 | 5 | 20 | 5 |
| 2005 | 859 | 426 | 60 | 10 | 60 | 10 | 10 | 5 | 20 | 5 |
| 2006 | 1,372 | 844 | 60 | 10 | 60 | 10 | 10 | 5 | 20 | 5 |
| 2007 | 784 | 707 | 60 | 10 | 60 | 10 | 10 | 5 | 20 | 5 |
| 2008 | 1,446 | 997 | 60 | 10 | 60 | 10 | 15 | 5 | 20 | 5 |
| 2009 | 2,882 | 1,080 | 60 | 10 | 60 | 10 | 15 | 5 | 20 | 5 |
| 2010 | 3,884 | 1,486 | 60 | 10 | 60 | 10 | 20 | 5 | 25 | 5 |
| 2011 | 3861 | 1407 | 60 | 10 | 60 | 10 | 20 | 5 | 25 | 5 |
| 2012 | 2708 | 1027 | 60 | 10 | 60 | 10 | 20 | 5 | 25 | 5 |
| $\begin{array}{r} \mathrm{M}(\min )= \\ \mathrm{M}(\max )= \end{array}$ | 0.020 |  | Return time (m)= |  | 1SW(min) | 6 MSW(min) | MSW (min) MSW (max) | 17 |  |  |

Russia (Kola Peninsula: White Sea Basin)

| Year | Catch (numbers) |  | Catch (numbers) previous year |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | 1SW | MSW | est | error | est | error | est | error | est | error |
| 1971 | 67,845 | 29,077 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1972 | 45,837 | 19,644 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1973 | 68,684 | 29,436 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1974 | 63,892 | 27,382 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1975 | 109,038 | 46,730 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1976 | 76,281 | 41,075 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1977 | 47,943 | 32,392 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1978 | 49,291 | 17,307 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1979 | 69,511 | 21,369 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1980 | 46,037 | 23,241 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1981 | 40,172 | 12,747 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1982 | 32,619 | 14,840 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1983 | 54,217 | 20,840 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1984 | 56,786 | 16,893 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1985 | 87,274 | 16,876 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1986 | 72,102 | 17,681 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 60 | 10 |
| 1987 | 79,639 | 12,501 | 0 | 0 | 3 | 2 | 3 | 2 | 50 | 10 | 50 | 10 |
| 1988 | 44,813 | 18,777 | 0 | 0 | 3 | 2 | 3 | 2 | 45 | 5 | 45 | 5 |
| 1989 | 53,293 | 11,448 | 0 | 0 | 7.5 | 2.5 | 7.5 | 2.5 | 45 | 5 | 45 | 5 |
| 1990 | 44,409 | 11,152 | 0 | 0 | 12.5 | 2.5 | 12.5 | 2.5 | 45 | 5 | 45 | 5 |
| 1991 | 31,978 | 6,263 | 0 | 0 | 17.5 | 2.5 | 17.5 | 2.5 | 35 | 5 | 35 | 5 |
| 1992 | 23,827 | 3,680 | 0 | 0 | 22.5 | 2.5 | 22.5 | 2.5 | 25 | 5 | 25 | 5 |
| 1993 | 20,987 | 5,552 | 0 | 0 | 25 | 5 | 25 | 5 | 25 | 5 | 25 | 5 |
| 1994 | 25,178 | 3,680 | 0 | 0 | 30 | 5 | 30 | 5 | 25 | 5 | 15 | 5 |
| 1995 | 19,381 | 2,847 | 0 | 0 | 35 | 5 | 35 | 5 | 25 | 5 | 15 | 5 |
| 1996 | 27,097 | 2,710 | 0 | 0 | 35 | 5 | 35 | 5 | 25 | 5 | 15 | 5 |
| 1997 | 27,695 | 2,085 | 0 | 0 | 35 | 5 | 35 | 5 | 25 | 5 | 15 | 5 |
| 1998 | 32,693 | 1,963 | 0 | 0 | 35 | 5 | 35 | 5 | 25 | 5 | 15 | 5 |
| 1999 | 22,330 | 2,841 | 0 | 0 | 35 | 5 | 35 | 5 | 25 | 5 | 15 | 5 |
| 2000 | 26,376 | 4,396 | 0 | 0 | 35 | 5 | 35 | 5 | 25 | 5 | 15 | 5 |
| 2001 | 20,483 | 3,959 | 0 | 0 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |
| 2002 | 19,174 | 3,937 | 0 | 0 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |
| 2003 | 15,687 | 3,734 | 0 | 0 | 35 | 5 | 25 | 5 | 15 | 5 | 15 | 5 |
| 2004 | 10,947 | 1,990 | 0 | 0 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |
| 2005 | 13,172 | 2,388 | 1212 | 878 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |
| 2006 | 15,004 | 2,071 | 3852 | 399 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |
| 2007 | 7,807 | 1,404 | 2264 | 852 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |
| 2008 | 8,447 | 4,711 | 3175 | 832 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |
| 2009 | 5,351 | 3,105 | 5130 | 1710 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |
| 2010 | 6,731 | 4,158 | 3684 | 1228 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |
| 2011 | 7363 | 4325 | 3082 | 1027 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |
| 2012 | 10398 | 1431 | 2267 | 756 | 35 | 5 | 35 | 5 | 15 | 5 | 15 | 5 |

## Russia (Pechora River)

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
| 1971 | 605 | 17,728 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1972 | 825 | 24,175 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1973 | 1,705 | 49,962 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1974 | 1,320 | 38,680 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1975 | 1,298 | 38,046 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1976 | 991 | 34,394 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1977 | 589 | 20,464 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1978 | 759 | 26,341 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1979 | 421 | 14,614 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1980 | 1,123 | 39,001 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1981 | 126 | 20,874 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1982 | 54 | 13,546 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1983 | 598 | 16,002 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1984 | 1,833 | 15,967 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1985 | 2,763 | 29,738 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1986 | 66 | 32,734 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1987 | 21 | 21,179 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |
| 1988 | 3,184 | 12,816 | 20 | 10 | 20 | 10 | 65 | 15 | 65 | 15 |

Russia (Pechora River) Continued.

|  | Estimated numbers of adult returns to fresh water |  | Input data for analysis of total adult returns to homewaters |  |  |  | Input data for spawner abundance analysis |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Marine unreported as \% of adult returns to freshwater |  |  |  | Freshwater unreported as \% of adult returns to freshwater |  |  |  |
|  |  |  | 1SW |  | MSW |  | 1SW |  | MSW |  |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
| 1989 | 24596 | 27404 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 1990 | 50 | 49950 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 1991 | 7975 | 47025 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 1992 | 550 | 54450 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 1993 | 68 | 67932 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 1994 | 3900 | 48100 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 1995 | 9280 | 70720 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 1996 | 8664 | 48336 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 1997 | 1440 | 38560 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 1998 | 780 | 59220 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 1999 | 2120 | 37880 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2000 | 84 | 83916 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2001 | 2244 | 41756 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2002 | 405 | 44595 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2003 | 1650 | 31350 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2004 | 6075 | 20925 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2005 | 2852 | 28148 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2006 | 1472 | 30528 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2007 | 817 | 42183 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2008 | 300 | 49700 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2009 | 1116 | 47385 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2010 | 1096 | 53704 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2011 | 2990 | 56810 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| 2012 | 4424 | 27176 | 10 | 5 | 10 | 5 | 65 | 15 | 65 | 15 |
| $\begin{array}{r} \mathrm{M}(\min )= \\ \mathrm{M}(\max )= \end{array}$ | 0.020 |  | Return time (m) |  | 1SW(min) | 7 8 | MSW(min) | 19 |  |  |

Sweden

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
| 1971 | 6,330 | 420 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1972 | 5,005 | 295 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1973 | 6,210 | 1,025 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1974 | 8,935 | 660 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1975 | 9,620 | 160 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1976 | 5,420 | 480 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1977 | 2,555 | 360 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1978 | 2,917 | 275 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1979 | 3,080 | 800 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1980 | 3,920 | 1,400 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1981 | 7,095 | 407 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1982 | 6,230 | 1,460 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1983 | 8,290 | 1,005 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1984 | 11,680 | 1,410 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1985 | 13,890 | 590 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1986 | 14,635 | 570 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1987 | 11,860 | 1,700 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1988 | 9,930 | 1,650 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1989 | 3,180 | 4,610 | 30 | 15 | 30 | 15 | 52.5 | 12.5 | 57.5 | 12.5 |
| 1990 | 7,430 | 3,135 | 15 | 10 | 15 | 10 | 45 | 15 | 50 | 15 |
| 1991 | 8,990 | 3,620 | 15 | 10 | 15 | 10 | 45 | 15 | 50 | 15 |
| 1992 | 9,850 | 4,655 | 15 | 10 | 15 | 10 | 45 | 15 | 50 | 15 |
| 1993 | 10,540 | 6,370 | 15 | 10 | 15 | 10 | 45 | 15 | 50 | 15 |
| 1994 | 8,035 | 4,660 | 15 | 10 | 15 | 10 | 45 | 15 | 50 | 15 |
| 1995 | 9,761 | 2,770 | 15 | 10 | 15 | 10 | 37.5 | 12.5 | 42.5 | 12.5 |
| 1996 | 6,008 | 3,542 | 15 | 10 | 15 | 10 | 37.5 | 12.5 | 42.5 | 12.5 |
| 1997 | 2,747 | 2,307 | 15 | 10 | 15 | 10 | 37.5 | 12.5 | 42.5 | 12.5 |
| 1998 | 2,421 | 1,702 | 15 | 10 | 15 | 10 | 37.5 | 12.5 | 42.5 | 12.5 |
| 1999 | 3,573 | 1,460 | 15 | 10 | 15 | 10 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2000 | 7,103 | 3,196 | 15 | 10 | 15 | 10 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2001 | 4,634 | 3,853 | 15 | 10 | 15 | 10 | 37.5 | 12.5 | 42.5 | 12.5 |
| 2002 | 4,733 | 2,826 | 15 | 10 | 15 | 10 | 38 | 13 | 43 | 13 |
| 2003 | 2,891 | 3,214 | 15 | 10 | 15 | 10 | 38 | 13 | 43 | 13 |
| 2004 | 2,494 | 2,330 | 15 | 10 | 15 | 10 | 38 | 13 | 43 | 13 |
| 2005 | 2,122 | 1,770 | 15 | 10 | 15 | 10 | 38 | 13 | 43 | 13 |
| 2006 | 2,585 | 1,772 | 15 | 10 | 15 | 10 | 38 | 13 | 43 | 13 |
| 2007 | 1,228 | 2,442 | 15 | 10 | 15 | 10 | 38 | 13 | 43 | 13 |
| 2008 | 1,197 | 2,752 | 13 | 8 | 13 | 8 | 28 | 13 | 33 | 13 |
| 2009 | 1,269 | 2,495 | 13 | 8 | 13 | 8 | 28 | 13 | 33 | 13 |
| 2010 | 2,109 | 3,066 | 13 | 8 | 13 | 8 | 28 | 13 | 33 | 13 |
| 2011 | 2726 | 5759 | 18 | 8 | 18 | 8 | 45 | 15 | 50 | 15 |
| 2012 | 1900 | 4826 | 13 | 8 | 8 | 5 | 28 | 13 | 33 | 13 |
| $\begin{array}{r} \mathrm{M}(\min )= \\ \mathrm{M}(\max )= \end{array}$ | 0.020 0.040 |  |  | me (m) $=$ | 1SW(min) 1SW(max) | 7 9 | MSW (min) MSW (max) | 16 18 |  |  |


|  | Total | Prop'n | NE coast catch |  |  |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1sw (\%) |  | Exp. rate MSW (\%) |  | NE unrep | Proportion Scottish in: |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | catch (numbers) | $\begin{aligned} & \text { 1SW in } \\ & \text { catch } \end{aligned}$ | Total | Dritt nets | T/J nets | Prop'n 1sw |  | error | est | error | est | error | est |  |  | Total | Drift | T/J |
| 1971 | 109,861 | 0.55 | 60,353 |  |  | 0.55 | 38.5 | 9.5 | 38.5 | 9.5 | 48 | 10 | 35 | 10 | 32.6 | 0.95 |  |  |
| 1972 | 108,074 | 0.42 | 51,681 |  |  | 0.42 | 39 | 10 | 39 | 10 | 47 | 10 | 35 | 10 | 32.6 | 0.95 |  |  |
| 1973 | 114,786 | 0.53 | 62,842 |  |  | 0.53 | 38.5 | 9.5 | 38.5 | 9.5 | 47 | 10 | 35 | 10 | 32.6 | 0.95 |  |  |
| 1974 | 104,325 | 0.65 | 52,756 |  |  | 0.65 | 39 | 10 | 39 | 10 | 47 | 10 | 35 | 10 | 32.6 | 0.95 |  |  |
| 1975 | 113,062 | 0.59 | 53,451 |  |  | 0.59 | 38.5 | 9.5 | 38.5 | 9.5 | 47 | 10 | 35 | 10 | 32.6 | 0.95 |  |  |
| 1976 | 54,294 | 0.64 | 15,701 |  |  | 0.64 | 37 | 9 | 37 | 9 | 48 | 10 | 36 | 10 | 32.6 | 0.94 |  |  |
| 1977 | 94,282 | 0.62 | 52,888 |  |  | 0.62 | 39 | 10 | 39 | 10 | 49 | 10 | 36 | 10 | 32.6 | 0.93 |  |  |
| 1978 | 93,125 | 0.69 | 51,630 |  |  | 0.69 | 38.5 | 9.5 | 38.5 | 9.5 | 49 | 10 | 36 | 10 | 32.6 | 0.92 |  |  |
| 1979 | 75,386 | 0.81 | 43,464 |  |  | 0.81 | 38.5 | 9.5 | 38.5 | 9.5 | 48 | 10 | 35 | 10 | 32.6 | 0.91 |  |  |
| 1980 | 90,218 | 0.55 | 45,780 |  |  | 0.55 | 39 | 10 | 39 | 10 | 48 | 10 | 36 | 10 | 32.6 | 0.9 |  |  |
| 1981 | 121,039 | 0.48 | 69,113 |  |  | 0.48 | 38.5 | 9.5 | 38.5 | 9.5 | 48 | 10 | 36 | 10 | 32.6 | 0.89 |  |  |
| 1982 | 80,289 | 0.67 | 50,167 |  |  | 0.67 | 38.5 | 9.5 | 38.5 | 9.5 | 48 | 10 | 36 | 10 | 32.6 | 0.88 |  |  |
| 1983 | 116,995 | 0.72 | 77,277 |  |  | 0.72 | 37 | 9 | 37 | 9 | 49 | 10 | 36 | 10 | 32.7 | 0.87 |  |  |
| 1984 | 94,271 | 0.74 | 59,295 |  |  | 0.74 | 36.5 | 9.5 | 36.5 | 9.5 | 49 | 10 | 36 | 10 | 32.6 | 0.86 |  |  |
| 1985 | 95,531 | 0.66 | 57,356 |  |  | 0.66 | 39 | 10 | 39 | 10 | 49 | 10 | 36 | 10 | 32.6 | 0.85 |  |  |
| 1986 | 110,794 | 0.62 | 63,425 |  |  | 0.62 | 37.5 | 9.5 | 37.5 | 9.5 | 49 | 10 | 36 | 10 | 32.6 | 0.84 |  |  |
| 1987 | 83,439 | 0.68 | 36,143 |  |  | 0.68 | 38.5 | 9.5 | 38.5 | 9.5 | 49 | 10 | 36 | 10 | 32.6 | 0.83 |  |  |
| 1988 | 110,163 | 0.69 |  | 47,465 | 3,384 | 0.69 | 40 | 10 | 40 | 10 | 48 | 10 | 36 | 10 | 32.6 |  | 0.82 | 0.5 |
| 1989 | 83,668 | 0.65 |  | 36,236 | 5,217 | 0.65 | 37 | 9 | 37 | 9 | 49 | 10 | 36 | 10 | 32.6 |  | 0.81 | 0.5 |
| 1990 | 86,676 | 0.52 |  | 48,219 | 3,311 | 0.52 | 37 | 9 | 37 | 9 | 49 | 10 | 36 | 10 | 31.6 |  | 0.8 | 0.5 |
| 1991 | 51,649 | 0.71 |  | 22,463 | 2,966 | 0.71 | 37.5 | 9.5 | 37.5 | 9.5 | 48 | 10 | 36 | 10 | 29.3 |  | 0.79 | 0.5 |
| 1992 | 44,586 | 0.77 |  | 17,574 | 2,570 | 0.77 | 40 | 10 | 40 | 10 | 48 | 10 | 36 | 10 | 26.9 |  | 0.78 | 0.5 |
| 1993 | 69,177 | 0.81 |  | 39,224 | 2,576 | 0.81 | 37.5 | 9.5 | 37.5 | 9.5 | 45 | 10 | 33 | 10 | 24.6 |  | 0.77 | 0.5 |
| 1994 | 88,121 | 0.77 |  | 41,298 | 5,256 | 0.77 | 24 | 6 | 24 | 6 | 45 | 10 | 33 | 10 | 22.4 |  | 0.76 | 0.5 |
| 1995 | 80,478 | 0.72 |  | 48,005 | 5,205 | 0.72 | 22.5 | 5.5 | 22.5 | 5.5 | 42 | 10 | 30 | 10 | 20.1 |  | 0.75 | 0.5 |
| 1996 | 46,696 | 0.65 |  | 15,172 | 3,409 | 0.65 | 20.5 | 5.5 | 20.5 | 5.5 | 41 | 10 | 30 | 10 | 17.9 |  | 0.75 | 0.5 |
| 1997 | 41,374 | 0.73 |  | 19,241 | 2,681 | 0.73 | 19 | 5 | 19 | 5 | 38 | 10 | 27 | 10 | 15.7 |  | 0.75 | 0.5 |
| 1998 | 36,917 | 0.82 |  | 17,328 | 937 | 0.82 | 19 | 5 | 19 | 5 | 35 | 10 | 25 | 10 | 15.7 |  | 0.75 | 0.5 |
| 1999 | 41,094 | 0.68 |  | 24,812 | 2,021 | 0.68 | 17.5 | 4.5 | 17.5 | 4.5 | 32 | 10 | 18 | 9 | 14.7 |  | 0.75 | 0.5 |
| 2000 | 60,953 | 0.79 |  | 40,059 | 3,295 | 0.79 | 15 | 4 | 15 | 4 | 32 | 10 | 15 | 8 | 6 |  | 0.75 | 0.5 |
| 2001 | 51,307 | 0.75 |  | 32,374 | 3,741 | 0.75 | 14.5 | 3.5 | 14.5 | 3.5 | 30 | 10 | 14 | 7 | 6 |  | 0.75 | 0.5 |
| 2002 | 45,669 | 0.76 |  | 27,685 | 3,295 | 0.76 | 15 | 4 | 15 | 4 | 30 | 10 | 14 | 7 | 6 |  | 0.75 | 0.5 |
| 2003 | 22,206 | 0.66 |  | 5,511 | 4,924 | 0.66 | 18 | 5 | 18 | 5 | 25 | 10 | 12 | 6 | 15 |  | 0.75 | 0.5 |
| 2004 | 30,559 | 0.81 |  | 5,921 | 5,096 | 0.81 | 18 | 5 | 18 | 5 | 28 | 10 | 12 | 6 | 15 |  | 0.75 | 0.5 |
| 2005 | 26,162 | 0.76 |  | 5,607 | 3,380 | 0.76 | 18 | 5 | 18 | 5 | 27 | 10 | 12 | 6 | 15 |  | 0.75 | 0.5 |
| 2006 | 22,056 | 0.78 |  | 4,040 | 3,526 | 0.78 | 18 | 5 | 18 | 5 | 25 | 10 | 11 | 5 | 15 |  | 0.75 | 0.5 |
| 2007 | 19,923 | 0.78 |  | 4,894 | 2,197 | 0.78 | 18 | 4 | 18 | 4 | 23 | 10 | 10 | 5 | 15 |  | 0.75 | 0.5 |
| 2008 | 19,036 | 0.76 |  | 3,649 | 2,592 | 0.76 | 18 | 4 | 18 | 4 | 23 | 10 | 10 | 5 | 15 |  | 0.75 | 0.5 |
| 2009 | 13,910 | 0.72 |  | 2,590 | 2,805 | 0.72 | 11 |  | 11 | 3 | 23 | 10 | 10 | 5 | 5 |  | 0.75 | 0.5 |
| 2010 | 32,695 | 0.78 |  | 12,214 | 7,768 |  |  |  |  |  |  |  |  |  |  |  | 0.75 | 0.5 |
| 2011 | 34575 | 0.57 |  | 15,517 | 8,631 | 0.78 | 11 | 3 | 11 | 3 | 23 | 10 | 10 | 5 | 5 |  | 0.75 | 0.5 |
| 2012 | 14699 | 0.51 |  | 3,805 | 3,471 | 0.57 | 11 | 3 | 11 | 3 | 25 | 10 | 12 | 6 | 5 |  | 0.75 | 0.5 |

## UK (N. Ireland)-Foyle Fisheries Area

| Year | Reported net catch (numbers) |  | Rod catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | 1SW | MSW | est | error | est | error | est | error | est | error |
| 1971 | 78,037 | 5,874 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1972 | 64,663 | 4,867 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1973 | 57,469 | 4,326 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1974 | 72,587 | 5,464 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1975 | 51,061 | 3,843 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1976 | 36,206 | 2,725 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1977 | 36,510 | 2,748 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1978 | 44,557 | 3,354 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1979 | 34,413 | 2,590 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1980 | 45,777 | 3,446 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1981 | 32,346 | 2,435 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1982 | 55,946 | 4,211 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1983 | 77,424 | 5,828 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1984 | 27,465 | 2,067 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1985 | 37,685 | 2,836 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1986 | 43,109 | 3,245 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1987 | 17,189 | 1,294 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 69 | 7 | 46 | 5 |
| 1988 | 43,974 | 3,310 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 64.5 | 6.5 | 36 | 4 |
| 1989 | 60,288 | 4,538 |  |  | 23.5 | 13.5 | 23.5 | 13.5 | 89 | 9 | 60 | 6 |
| 1990 | 39,875 | 3,001 |  |  | 13.5 | 3.5 | 13.5 | 3.5 | 62 | 6 | 38 | 4 |
| 1991 | 21,709 | 1,634 |  |  | 13.5 | 3.5 | 13.5 | 3.5 | 64.5 | 6.5 | 43 | 4 |
| 1992 | 39,299 | 2,958 |  |  | 16.5 | 6.5 | 16.5 | 6.5 | 56 | 6 | 33 | 3 |
| 1993 | 35,366 | 2,662 |  |  | 13.5 | 3.5 | 13.5 | 3.5 | 41 | 4 | 12 | 1 |
| 1994 | 36,144 | 2,720 |  |  | 19 | 9 | 19 | 9 | 70 | 7 | 40 | 4 |
| 1995 | 33,398 | 2,514 |  |  | 13.5 | 3.5 | 13.5 | 3.5 | 67 | 7 | 42 | 4 |
| 1996 | 28,406 | 2,138 |  |  | 15 | 5 | 15 | 5 | 57 | 10 | 34 | 10 |
| 1997 | 40,886 | 3,077 |  |  | 10 | 5 | 10 | 5 | 60 | 10 | 34 | 10 |
| 1998 | 37,154 | 2,797 |  |  | 10 | 5 | 10 | 5 | 25 | 5 | 23 | 8 |
| 1999 | 21,660 | 1,630 |  |  | 10 | 5 | 10 | 5 | 63 | 5 | 33 | 8 |
| 2000 | 30,385 | 2,287 |  |  | 10 | 5 | 10 | 5 | 58 | 5 | 33 | 8 |
| 2001 | 21,368 | 1,608 |  |  | 5 | 5 | 5 | 5 | 50 | 5 | 30 | 5 |
| 2002 | 37,914 | 2,854 | 9,163 | 690 | 3 | 2 | 3 | 3 | 15 | 3 | 15 | 3 |
| 2003 | 30,441 | 2,291 | 4,576 | 344 | 1 | 0 | 1 | 0 | 15 | 3 | 15 | 3 |
| 2004 | 20,730 | 1,560 | 4,570 | 344 | 1 | 0 | 1 | 0 | 15 | 3 | 15 | 3 |
| 2005 | 23,746 | 1,787 | 7,079 | 533 | 1 | 0 | 1 | 0 | 15 | 3 | 15 | 3 |
| 2006 | 11,324 | 852 | 4,886 | 368 | 1 | 0 | 1 | 0 | 15 | 3 | 15 | 3 |
| 2007 | 5,050 | 322 | 9,530 | 608 | 1 | 1 | 1 | 1 | 15 | 3 | 15 | 3 |
| 2008 | 3,880 | 292 | 4,755 | 304 | 1 | 0 | 1 | 0 | 15 | 3 | 15 | 3 |
| 2009 | 1,743 | 194 | 3,640 | 405 | 1 | 0 | 1 | 0 | 15 | 3 | 15 | 3 |
| 2010 | 0 | 0 | 4,257 | 473 | 1 | 0 | 1 | 0 | 15 | 3 | 15 | 3 |
| 2011 | 0 | 0 | 3770 | 1256 | 1 | 0 | 1 | 0 | 15 | 5 | 15 | 5 |
| 2012 | 0 | 0 | 4687 | 1563 | 1 | 0 | 1 | 0 | 13 | 8 | 13 | 8 |
| $M(\min )=$ | 0.020 |  |  |  | Return time ( m ) $=$ |  | 1SW(min) | $7 \quad \mathrm{MSW}$ (min) |  | 16 |  |  |
| $\mathrm{M}(\max )=$ | 0.040 |  |  |  |  |  | 1SW(max) | 9 | MSW(max) | 18 |  |  |

Note Assessment based on net catches for the period 1971 to 2001 and rod catches thereafter

## UK ( N. Ireland)-DCAL area

| Year | Reported net catch (numbers) |  | Rod catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | 1SW | MSW | est | error | est | error | est | error | est | error |
| 1971 | 35,506 | 2,673 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1972 | 34,550 | 2,601 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1973 | 29,229 | 2,200 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1974 | 22,307 | 1,679 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1975 | 26,701 | 2,010 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1976 | 17,886 | 1,346 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1977 | 16,778 | 1,263 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1978 | 24,857 | 1,871 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1979 | 14,323 | 1,078 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1980 | 15,967 | 1,202 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1981 | 15,994 | 1,204 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1982 | 14,068 | 1,059 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1983 | 20,845 | 1,569 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1984 | 11,109 | 836 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1985 | 12,369 | 931 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1986 | 13,160 | 991 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 80 | 5 | 50 | 5 |
| 1987 | 9,240 | 695 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 69 | 7 | 46 | 5 |
| 1988 | 14,320 | 1,078 |  |  | 21.5 | 11.5 | 21.5 | 11.5 | 64.5 | 6.5 | 36 | 4 |
| 1989 | 15,081 | 1,135 |  |  | 23.5 | 13.5 | 23.5 | 13.5 | 89 | 9 | 60 | 6 |
| 1990 | 9,499 | 715 |  |  | 13.5 | 3.5 | 13.5 | 3.5 | 62 | 6 | 38 | 4 |
| 1991 | 6,987 | 526 |  |  | 13.5 | 3.5 | 13.5 | 3.5 | 64.5 | 6.5 | 43 | 4 |
| 1992 | 9,346 | 703 |  |  | 16.5 | 6.5 | 16.5 | 6.5 | 56 | 6 | 33 | 3 |
| 1993 | 7,906 | 595 |  |  | 13.5 | 3.5 | 13.5 | 3.5 | 41 | 4 | 12 | 1 |
| 1994 | 11,206 | 843 |  |  | 19 | 9 | 19 | 9 | 70 | 7 | 40 | 4 |
| 1995 | 11,637 | 876 |  |  | 13.5 | 3.5 | 13.5 | 3.5 | 67 | 7 | 42 | 4 |
| 1996 | 10,383 | 781 |  |  | 15 | 5 | 15 | 5 | 57 | 10 | 34 | 10 |
| 1997 | 10,479 | 789 |  |  | 10 | 5 | 10 | 5 | 60 | 10 | 34 | 10 |
| 1998 | 9,375 | 706 |  |  | 10 | 5 | 10 | 5 | 25 | 5 | 23 | 8 |
| 1999 | 9,011 | 678 |  |  | 10 | 5 | 10 | 5 | 63 | 5 | 33 | 8 |
| 2000 | 10,598 | 798 |  |  | 10 | 5 | 10 | 5 | 58 | 5 | 33 | 8 |
| 2001 | 8,104 | 610 |  |  | 5 | 5 | 5 | 5 | 50 | 5 | 30 | 5 |
| 2002 | 3,315 | 249 | 2,218 | 167 | 3 | 3 | 3 | 3 | 14 | 9 | 14 | 9 |
| 2003 | 2,236 | 168 | 1,884 | 141 | 3 | 3 | 3 | 3 | 12 | 7 | 12 | 7 |
| 2004 | 2,411 | 181 | 3,053 | 230 | 1 | 1 | 1 | 1 | 18 | 10 | 18 | 10 |
| 2005 | 3,012 | 227 | 1,791 | 135 | 1 | 1 | 1 | 1 | 12 | 7 | 12 | 7 |
| 2006 | 2,288 | 172 | 1,289 | 97 | 1 | 1 | 1 | 1 | 12 | 8 | 12 | 8 |
| 2007 | 2,533 | 162 | 2,427 | 155 | 1 | 1 | 1 | 1 | 11 | 4 | 11 | 4 |
| 2008 | 1,825 | 116 | 2,444 | 156 | 1 | 0 | 1 | 0 | 14 | 7 | 14 | 7 |
| 2009 | 1,383 | 154 | 1,457 | 162 | 1 | 0 | 1 | 0 | 10 | 3 | 10 | 3 |
| 2010 | 1,723 | 191 | 1,327 | 147 | 1 | 0 | 1 | 0 | 15 | 3 | 15 | 3 |
| 2011 | 857 | 285 | 1132 | 378 | 1 | 0 | 1 | 0 | 15 | 5 | 15 | 5 |
| 2012 | 15 | 5 | 427 | 143 | 1 | 0 | 1 | 0 |  |  |  |  |
| $M(\min )=$ | 0.020 |  |  |  | Return time (m)= |  | 1SW(min) | 7 | MSW(min) | 16 |  |  |
| $\mathrm{M}(\max )=$ | 0.040 |  |  |  |  |  | 1SW(max) | 9 | MSW(max) | 18 |  |  |

Note Assessment based on net catches for the period 1971 to 2001 and rod catches thereafter

UK (Scotland)-East

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
| 1971 | 216,873 | 135,530 | 25 | 10 | 25 | 10 | 75 | 13 | 50 | 10 |
| 1972 | 220,106 | 183,875 | 25 | 10 | 25 | 10 | 77 | 13 | 51 | 10 |
| 1973 | 259,773 | 204,826 | 25 | 10 | 25 | 10 | 75 | 12 | 50 | 10 |
| 1974 | 245,424 | 158,959 | 25 | 10 | 25 | 10 | 82 | 14 | 56 | 11 |
| 1975 | 181,940 | 180,828 | 25 | 10 | 25 | 10 | 80 | 13 | 55 | 11 |
| 1976 | 150,069 | 92,179 | 25 | 10 | 25 | 10 | 77 | 13 | 51 | 10 |
| 1977 | 154,306 | 118,645 | 25 | 10 | 25 | 10 | 81 | 14 | 56 | 11 |
| 1978 | 158,859 | 139,763 | 25 | 10 | 25 | 10 | 76 | 13 | 51 | 10 |
| 1979 | 160,796 | 116,559 | 25 | 10 | 25 | 10 | 78 | 13 | 54 | 11 |
| 1980 | 101,665 | 155,646 | 17.5 | 7.5 | 17.5 | 7.5 | 77 | 13 | 52 | 10 |
| 1981 | 129,690 | 156,683 | 17.5 | 7.5 | 17.5 | 7.5 | 76 | 13 | 51 | 10 |
| 1982 | 175,374 | 113,198 | 17.5 | 7.5 | 17.5 | 7.5 | 71 | 12 | 45 | 9 |
| 1983 | 170,843 | 126,104 | 17.5 | 7.5 | 17.5 | 7.5 | 77 | 13 | 49 | 10 |
| 1984 | 175,675 | 90,829 | 17.5 | 7.5 | 17.5 | 7.5 | 70 | 12 | 44 | 9 |
| 1985 | 133,119 | 95,044 | 17.5 | 7.5 | 17.5 | 7.5 | 62 | 10 | 39 | 8 |
| 1986 | 180,292 | 128,654 | 17.5 | 7.5 | 17.5 | 7.5 | 59 | 10 | 38 | 8 |
| 1987 | 139,252 | 88,519 | 17.5 | 7.5 | 17.5 | 7.5 | 65 | 11 | 41 | 8 |
| 1988 | 118,614 | 91,151 | 17.5 | 7.5 | 17.5 | 7.5 | 40 | 7 | 29 | 6 |
| 1989 | 143,049 | 85,385 | 10 | 5 | 10 | 5 | 38 | 6 | 28 | 6 |
| 1990 | 63,318 | 73,971 | 10 | 5 | 10 | 5 | 40 | 7 | 29 | 6 |
| 1991 | 53,860 | 53,693 | 10 | 5 | 10 | 5 | 37 | 6 | 27 | 5 |
| 1992 | 79,883 | 67,968 | 10 | 5 | 10 | 5 | 32 | 5 | 26 | 5 |
| 1993 | 73,396 | 60,496 | 10 | 5 | 10 | 5 | 35 | 6 | 27 | 5 |
| 1994 | 80,429 | 72,758 | 10 | 5 | 10 | 5 | 33 | 6 | 26 | 5 |
| 1995 | 72,973 | 69,051 | 10 | 5 | 10 | 5 | 31 | 5 | 25 | 5 |
| 1996 | 56,627 | 50,365 | 10 | 5 | 10 | 5 | 29 | 5 | 24 | 5 |
| 1997 | 37,448 | 34,850 | 10 | 5 | 10 | 5 | 31 | 5 | 25 | 5 |
| 1998 | 44,952 | 32,231 | 10 | 5 | 10 | 5 | 24 | 4 | 23 | 5 |
| 1999 | 20,907 | 27,011 | 10 | 5 | 10 | 5 | 25 | 4 | 23 | 5 |
| 2000 | 36,871 | 31,280 | 10 | 5 | 10 | 5 | 22 | 4 | 22 | 4 |
| 2001 | 36,646 | 30,470 | 10 | 5 | 10 | 5 | 20 | 3 | 22 | 5 |
| 2002 | 26,616 | 21,740 | 10 | 5 | 10 | 5 | 19 | 3 | 21 | 4 |
| 2003 | 25,871 | 24,270 | 10 | 5 | 10 | 5 | 17 | 3 | 19 | 4 |
| 2004 | 31,667 | 30,773 | 10 | 5 | 10 | 5 | 17 | 3 | 19 | 4 |
| 2005 | 31,597 | 23,676 | 10 | 5 | 10 | 5 | 17 | 3 | 19 | 4 |
| 2006 | 30,739 | 22,954 | 10 | 5 | 10 | 5 | 15 | 3 | 17 | 4 |
| 2007 | 26,015 | 19,444 | 10 | 5 | 10 | 5 | 14 | 3 | 15 | 4 |
| 2008 | 18,586 | 20,757 | 10 | 5 | 10 | 5 | 11 | 3 | 14 | 4 |
| 2009 | 14,863 | 15,042 | 10 | 5 | 10 | 5 | 10 | 3 | 13 | 4 |
| 2010 | 28,252 | 22,908 | 10 | 5 | 10 | 5 | 10 | 3 | 13 | 4 |
| 2011 | 12485 | 24213 | 10 | 5 | 10 | 5 | 9 | 3 | 13 | 4 |
| 2012 | 17026 | 17046 | 10 | 5 | 10 | 5 | 8 | 3 | 12 | 4 |
| $\begin{array}{r} \mathrm{M}(\min )= \\ \mathrm{M}(\max )= \end{array}$ | 0.020 0.040 |  | Return time (m)= |  | $\begin{aligned} & \text { 1SW(min) } \\ & \text { 1SW(max) } \end{aligned}$ | 7 MSW(min) |  | 17 18 |  |  |

## UK (Scotland)-West

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |
| 1971 | 45,287 | 26,071 | 35 | 10 | 35 | 10 | 38 | 6 | 25 | 5 |
| 1972 | 31,358 | 34,148 | 35 | 10 | 35 | 10 | 38 | 6 | 26 | 5 |
| 1973 | 33,317 | 33,094 | 35 | 10 | 35 | 10 | 37 | 6 | 25 | 5 |
| 1974 | 43,992 | 29,369 | 35 | 10 | 35 | 10 | 41 | 7 | 28 | 6 |
| 1975 | 40,424 | 27,145 | 35 | 10 | 35 | 10 | 40 | 7 | 28 | 6 |
| 1976 | 38,409 | 22,367 | 35 | 10 | 35 | 10 | 38 | 6 | 25 | 5 |
| 1977 | 39,952 | 20,335 | 35 | 10 | 35 | 10 | 41 | 7 | 28 | 6 |
| 1978 | 45,611 | 23,191 | 35 | 10 | 35 | 10 | 38 | 6 | 25 | 5 |
| 1979 | 26,440 | 15,950 | 35 | 10 | 35 | 10 | 39 | 7 | 27 | 5 |
| 1980 | 19,776 | 16,942 | 27.5 | 7.5 | 27.5 | 7.5 | 38 | 6 | 26 | 5 |
| 1981 | 21,048 | 18,038 | 27.5 | 7.5 | 27.5 | 7.5 | 38 | 6 | 26 | 5 |
| 1982 | 32,687 | 15,044 | 27.5 | 7.5 | 27.5 | 7.5 | 36 | 6 | 23 | 5 |
| 1983 | 38,774 | 19,857 | 27.5 | 7.5 | 27.5 | 7.5 | 39 | 6 | 25 | 5 |
| 1984 | 37,404 | 16,384 | 27.5 | 7.5 | 27.5 | 7.5 | 35 | 6 | 22 | 4 |
| 1985 | 24,861 | 19,571 | 27.5 | 7.5 | 27.5 | 7.5 | 31 | 5 | 19 | 4 |
| 1986 | 22,546 | 19,543 | 27.5 | 7.5 | 27.5 | 7.5 | 30 | 5 | 19 | 4 |
| 1987 | 25,533 | 15,475 | 27.5 | 7.5 | 27.5 | 7.5 | 32 | 5 | 20 | 4 |
| 1988 | 30,484 | 21,011 | 27.5 | 7.5 | 27.5 | 7.5 | 20 | 3 | 15 | 3 |
| 1989 | 31,892 | 18,501 | 20 | 5 | 20 | 5 | 19 | 3 | 14 | 3 |
| 1990 | 17,776 | 13,953 | 20 | 5 | 20 | 5 | 20 | 3 | 14 | 3 |
| 1991 | 19,748 | 11,500 | 20 | 5 | 20 | 5 | 18 | 3 | 14 | 3 |
| 1992 | 21,793 | 14,873 | 20 | 5 | 20 | 5 | 16 | 3 | 13 | 3 |
| 1993 | 21,121 | 11,230 | 20 | 5 | 20 | 5 | 18 | 3 | 13 | 3 |
| 1994 | 18,234 | 12,304 | 20 | 5 | 20 | 5 | 17 | 3 | 13 | 3 |
| 1995 | 16,831 | 9,137 | 20 | 5 | 20 | 5 | 15 | 3 | 13 | 3 |
| 1996 | 9,537 | 7,463 | 20 | 5 | 20 | 5 | 14 | 2 | 12 | 2 |
| 1997 | 9,059 | 5,504 | 20 | 5 | 20 | 5 | 15 | 3 | 13 | 3 |
| 1998 | 8,369 | 6,150 | 20 | 5 | 20 | 5 | 12 | 2 | 11 | 2 |
| 1999 | 4,147 | 3,587 | 20 | 5 | 20 | 5 | 12 | 2 | 12 | 2 |
| 2000 | 6,974 | 5,301 | 20 | 5 | 20 | 5 | 11 | 2 | 11 | 2 |
| 2001 | 5,603 | 4,191 | 20 | 5 | 20 | 5 | 10 | 2 | 11 | 2 |
| 2002 | 4,691 | 4,548 | 20 | 5 | 20 | 5 | 10 | 2 | 11 | 2 |
| 2003 | 3,536 | 3,061 | 20 | 5 | 20 | 5 | 5 | 1 | 5 | 1 |
| 2004 | 5,836 | 6,024 | 20 | 5 | 20 | 5 | 7 | 1 | 8 | 2 |
| 2005 | 7,428 | 4,913 | 20 | 5 | 20 | 5 | 7 | 1 | 8 | 2 |
| 2006 | 5,767 | 4,403 | 20 | 5 | 20 | 5 | 7 | 1 | 8 | 2 |
| 2007 | 6,178 | 4,470 | 20 | 5 | 20 | 5 | 7 | 1 | 8 | 2 |
| 2008 | 4,740 | 4,853 | 20 | 5 | 20 | 5 | 7 | 1 | 8 | 2 |
| 2009 | 3,250 | 4,095 | 20 | 5 | 20 | 5 | 6 | 1 | 7 | 2 |
| 2010 | 5,107 | 4,052 | 20 | 5 | 20 | 5 | 6 | 1 | 7 | 2 |
| 2011 | 3206 | 4246 | 20 | 5 | 20 | 5 | 6 | 1 | 6 | 2 |
| 2012 | 3388 | 3537 | 20 | 5 | 20 | 5 | 5 | 1 | 5 | 2 |
| $\begin{aligned} M(\min ) & = \\ M(\max ) & = \end{aligned}$ | 0.020 0.040 |  | Return time ( m ) = |  | 1SW(min) | 7 | MSW(min) | 16 |  |  |

Faroes

| Year | Catch (numbers) |  | Unreported as \% of total 1SW |  | Unreported as \% of total 1SW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  | Prop'n wild | Stock composition |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | est | error | est | error | est | error | est | error |  | Country | 1SW | MSW |
|  |  |  |  |  |  |  |  |  |  |  |  | NNEAC |  |  |
| 1971 | 2,620 | 105,796 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 1 | Finland | 0.059 | 0.050 |
| 1972 | 2,754 | 111,187 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 1 | Iceland-NE | 0.016 | 0.011 |
| 1973 | 3,121 | 126,012 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 1 | Norway | 0.290 | 0.295 |
| 1974 | 2,186 | 88,276 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 1 | Russia | 0.116 | 0.163 |
| 1975 | 2,798 | 112,984 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 1 | Sweden | 0.019 | 0.016 |
| 1976 | 1,830 | 73,900 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 1 |  |  |  |
| 1977 | 1,291 | 52,112 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 1 | SNEAC |  |  |
| 1978 | 974 | 39,309 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 1 | France | 0.018 | 0.005 |
| 1979 | 1,736 | 70,082 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 1 | Iceland-SW | 0.025 | 0.007 |
| 1980 | 4,523 | 182,616 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 1 | Ireland | 0.173 | 0.043 |
| 1981 | 7,443 | 300,542 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.98 | UK(England | 0.044 | 0.034 |
| 1982 | 6,859 | 276,957 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.98 | UK(N. Ireland | 0.046 | 0.014 |
| 1983 | 15,861 | 215,349 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.98 | UK(Scotland | 0.195 | 0.337 |
| 1984 | 5,534 | 138,227 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.96 |  |  |  |
| 1985 | 378 | 158,103 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.92 | Other | 0.000 | 0.025 |
| 1986 | 1,979 | 180,934 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.96 |  |  |  |
| 1987 | 90 | 166,244 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.97 | Total | 1 | 1 |
| 1988 | 8,637 | 87,629 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.92 |  |  |  |
| 1989 | 1,788 | 121,965 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.82 |  |  |  |
| 1990 | 1,989 | 140,054 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.54 |  |  |  |
| 1991 | 943 | 84,935 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.54 |  |  |  |
| 1992 | 68 | 35,700 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.62 |  |  |  |
| 1993 | 6 | 30,023 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.69 |  |  |  |
| 1994 | 15 | 31,672 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.72 |  |  |  |
| 1995 | 18 | 34,662 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 1996 | 101 | 28,381 | 10 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.75 |  |  |  |
| 1997 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 1998 | 339 | 1,424 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 1999 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2000 | 225 | 1,765 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2001 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2002 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2003 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2004 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2005 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2006 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2007 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2008 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2009 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2010 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2011 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
| 2012 | 0 | 0 | 15 | 5 | 0 | 0 | 100 | 100 | 100 | 100 | 0.8 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\mathrm{M}(\mathrm{min})=$ | 0.020 |  | Ret | ime (m) $=$ | 1SW(min) | 0 | MSW(min) | 13 |  |  |  |  |  |  |
| $\mathrm{M}(\max )=$ | 0.040 |  |  |  | 1sW(max) | 1 | MSW(max) | 14 |  |  |  |  |  |  |
|  |  |  | Proportion of 1SW returning as grilse $=$ |  |  |  | min | 0.730 |  |  |  |  |  |  |
|  |  |  |  |  |  |  | max | 0.830 |  |  |  |  |  |  |

West Greenland


Annex 6: Input data for run-reconstruction of Atlantic salmon in the NAC area used to do the run-reconstruction and estimates of returns and spawners by size group and age group for North America
6.i. Input data for the fishery at West Greenland used in the run reconstruction model.

| $\begin{aligned} & \text { Year of the } \\ & \text { fishery } \\ & \hline \end{aligned}$ | Reported harvest (t) | Unreported harvest estimate (t) | Mean weight of salmon (all age groups, kg) | Genetic samples |  | Prop. NAC minimum | Prop. NAC maximum | Prop. 1SW salmon in the |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | NAC origin | NEAC origin |  |  | NAC | NEAC |
| 1971 | 2689 | 0 | 3.14 | 0 | 0 | 0.28 | 0.40 | 0.945 | 0.964 |
| 1972 | 2113 | 0 | 3.44 | 0 | 0 | 0.34 | 0.37 | 0.945 | 0.964 |
| 1973 | 2341 | 0 | 4.18 | 0 | 0 | 0.39 | 0.59 | 0.945 | 0.964 |
| 1974 | 1917 | 0 | 3.58 | 0 | 0 | 0.39 | 0.46 | 0.945 | 0.964 |
| 1975 | 2030 | 0 | 3.12 | 0 | 0 | 0.40 | 0.48 | 0.945 | 0.964 |
| 1976 | 1175 | 0 | 3.04 | 0 | 0 | 0.38 | 0.48 | 0.945 | 0.964 |
| 1977 | 1420 | 0 | 3.21 | 0 | 0 | 0.38 | 0.57 | 0.945 | 0.964 |
| 1978 | 984 | 0 | 3.35 | 0 | 0 | 0.47 | 0.57 | 0.945 | 0.964 |
| 1979 | 1395 | 0 | 3.34 | 0 | 0 | 0.48 | 0.52 | 0.945 | 0.964 |
| 1980 | 1194 | 0 | 3.22 | 0 | 0 | 0.45 | 0.51 | 0.945 | 0.964 |
| 1981 | 1264 | 0 | 3.17 | 0 | 0 | 0.58 | 0.61 | 0.945 | 0.964 |
| 1982 | 1077 | 0 | 3.11 | 0 | 0 | 0.60 | 0.64 | 0.945 | 0.964 |
| 1983 | 310 | 0 | 3.10 | 0 | 0 | 0.38 | 0.41 | 0.945 | 0.964 |
| 1984 | 297 | 0 | 3.11 | 0 | 0 | 0.47 | 0.53 | 0.945 | 0.964 |
| 1985 | 864 | 0 | 2.87 | 0 | 0 | 0.46 | 0.53 | 0.925 | 0.950 |
| 1986 | 960 | 0 | 3.03 | 0 | 0 | 0.48 | 0.66 | 0.951 | 0.975 |
| 1987 | 966 | 0 | 3.16 | 0 | 0 | 0.54 | 0.63 | 0.963 | 0.980 |
| 1988 | 893 | 0 | 3.18 | 0 | 0 | 0.38 | 0.49 | 0.967 | 0.981 |
| 1989 | 337 | 0 | 2.87 | 0 | 0 | 0.52 | 0.60 | 0.923 | 0.955 |
| 1990 | 274 | 0 | 2.69 | 0 | 0 | 0.70 | 0.79 | 0.957 | 0.963 |
| 1991 | 472 | 0 | 2.65 | 0 | 0 | 0.61 | 0.69 | 0.956 | 0.934 |
| 1992 | 237 | 0 | 2.81 | 0 | 0 | 0.50 | 0.57 | 0.919 | 0.975 |
| 1993 | 0 | 12 | 2.73 | 0 | 0 | 0.50 | 0.76 | 0.95 | 0.96 |
| 1994 | 0 | 12 | 2.73 | 0 | 0 | 0.50 | 0.76 | 0.95 | 0.96 |
| 1995 | 83 | 20 | 2.56 | 0 | 0 | 0.65 | 0.72 | 0.968 | 0.973 |
| 1996 | 92 | 20 | 2.88 | 0 | 0 | 0.71 | 0.76 | 0.941 | 0.961 |
| 1997 | 58 | 5 | 2.71 | 0 | 0 | 0.75 | 0.84 | 0.982 | 0.993 |
| 1998 | 11 | 11 | 2.78 | 0 | 0 | 0.73 | 0.84 | 0.968 | 0.994 |
| 1999 | 19 | 12.5 | 3.08 | 0 | 0 | 0.84 | 0.97 | 0.968 | 1.000 |
| 2000 | 21 | 10 | 2.57 | 344 | 146 | 0 | 0 | 0.974 | 1.000 |
| 2001 | 43 | 10 | 3.00 | - 1 | 1 | 0.67 | 0.71 | 0.982 | 0.978 |
| 2002 | 9.8 | 10 | 2.90 | 338 | 163 | 0 | 0 | 0.973 | 1.000 |
| 2003 | 12.3 | 10 | 3.04 | 1212 | 567 | 0 | 0 | 0.967 | 0.989 |
| 2004 | 17.2 | 10 | 3.18 | 1192 | 447 | 0 | 0 | 0.970 | 0.970 |
| 2005 | 17.3 | 10 | 3.31 | 585 | 182 | 0 | 0 | 0.924 | 0.967 |
| 2006 | 23.0 | 10 | 3.24 | 857 | 326 | 0 | 0 | 0.930 | 0.988 |
| 2007 | 24.8 | 10 | 2.98 | 917 | 206 | , | 0 | 0.965 | 0.956 |
| 2008 | 28.6 | 10 | 3.08 | 1593 | 260 | 0 | 0 | 0.974 | 0.988 |
| 2009 | 28.0 | 10 | 3.50 | 1483 | 138 | 0 | 0 | 0.934 | 0.894 |
| 2010 | 43.1 | 10 | 3.42 | 991 | 249 |  | 0 | 0.982 | 0.975 |
| 2011 | 27.4 | 10 | 3.69 | 888 | 72 | 0 | 0 | 0.939 | 0.831 |
| 2012 | 34.5 | 10 | 3.44 | 1121 | 252 | 0 | 0 | 0.932 | - 0.980 |
| Winbugs labels | WGHari] | WGUnHari] | WGMeanWt[] | WGSampleNAC | WGSampleNEA | AWGPropNAd | WGPropNACM | WGProp1SWN AC[] | $\begin{aligned} & \text { WGProp1SW } \\ & \text { NEAC] } \end{aligned}$ |

6.ii. Input data for sea fisheries on large salmon and small salmon from Newfoundland and Labrador used in the run reconstruction model. Labrador represents harvests from Labrador in aboriginal fisheries for food, social and ceremonial purposes and the resident food fishery beginning in 1998.

|  | Catches of large salmon |  |  | Catches of small salmon |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year of the fishery | SFA 1 to 7 | SFA 8 to 14A | FSC Labrador | SFA 1 to 7 | SFA 8 to 14A | FSC Labrador |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 199176 | 0 | 0 | 158896 | 70936 | 0 |
| 1972 | 144496 | 42861 | 0 | 143232 | 111141 | 0 |
| 1973 | 227779 | 43627 | 0 | 188725 | 176907 | 0 |
| 1974 | 196726 | 85714 | 0 | 192195 | 153278 | 0 |
| 1975 | 215025 | 72814 | 0 | 302348 | 91935 | 0 |
| 1976 | 210858 | 95714 | 0 | 221766 | 118779 | 0 |
| 1977 | 231393 | 63449 | 0 | 220093 | 57472 | 0 |
| 1978 | 155546 | 37653 | 0 | 102403 | 38180 | 0 |
| 1979 | 82174 | 29122 | 0 | 186558 | 62622 | 0 |
| 1980 | 211896 | 54307 | 0 | 290127 | 94291 | 0 |
| 1981 | 211006 | 38663 | 0 | 288902 | 60668 | 0 |
| 1982 | 129319 | 35055 | 0 | 222894 | 77017 | 0 |
| 1983 | 108430 | 28215 | 0 | 166033 | 55683 | 0 |
| 1984 | 87742 | 15135 | 0 | 123774 | 52813 | 0 |
| 1985 | 70970 | 24383 | 0 | 178719 | 79275 | 0 |
| 1986 | 107561 | 22036 | 0 | 222671 | 91912 | 0 |
| 1987 | 146242 | 19241 | 0 | 281762 | 82401 | 0 |
| 1988 | 86047 | 14763 | 0 | 198484 | 74620 | 0 |
| 1989 | 85319 | 15577 | 0 | 172861 | 60884 | 0 |
| 1990 | 59334 | 11639 | 0 | 104788 | 46053 | 0 |
| 1991 | 39257 | 10259 | 0 | 89099 | 42721 | 0 |
| 1992 | 32341 | 0 | 0 | 24249 | 0 | 0 |
| 1993 | 17096 | 0 | 0 | 17074 | 0 | 0 |
| 1994 | 15377 | 0 | 0 | 8640 | 0 | 0 |
| 1995 | 11176 | 0 | 0 | 7980 | 0 | 0 |
| 1996 | 7272 | 0 | 0 | 7849 | 0 | 0 |
| 1997 | 6943 | 0 | 0 | 9753 | 0 | 0 |
| 1998 | 0 | 0 | 2269 | 0 | 0 | 2988 |
| 1999 | 0 | 0 | 1084 | 0 | 0 | 2739 |
| 2000 | 0 | 0 | 1352 | 0 | 0 | 5323 |
| 2001 | 0 | 0 | 1721 | 0 | 0 | 4789 |
| 2002 | 0 | 0 | 1389 | 0 | 0 | 5806 |
| 2003 | 0 | 0 | 2175 | 0 | 0 | 6477 |
| 2004 | 0 | 0 | 3696 | 0 | 0 | 8385 |
| 2005 | 0 | 0 | 2817 | 0 | 0 | 10436 |
| 2006 | 0 | 0 | 3090 | 0 | 0 | 10377 |
| 2007 | 0 | 0 | 2652 | 0 | 0 | 9208 |
| 2008 | 0 | 0 | 3909 | 0 | 0 | 9834 |
| 2009 | 0 | 0 | 3344 | 0 | 0 | 7988 |
| 2010 | 0 | 0 | 3725 | 0 | 0 | 9867 |
| 2011 | 0 | 0 | 4447 | 0 | 0 | 11142 |
| 2012 | 0 | 0 | 4230 | 0 | 0 | 9988 |
|  |  |  |  |  |  |  |
| Winbugs labels | NIg_LBandNF1to7] | NIg_NF8to14a[] | Nlg_LBFSC[] | Nsm_LBandNF1to7] | Nsm_NF8to14a[] | Nsm_LBFSC] |

6.iii. Input data for sea fisheries on large salmon and small salmon from St-Pierre and Miquelon used in the run reconstruction model.

| Year of the fishery | Reported harvest (kg) | Catch in number (all sizes) | Catch in number (large) | Catch in number (small) |
| :---: | :---: | :---: | :---: | :---: |
| 1970 | 0 | 0 | 0 | 0 |
| 1971 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 |
| 1976 | 3000 | 1331 | 333 | 998 |
| 1977 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 |
| 1983 | 3000 | 1331 | 333 | 998 |
| 1984 | 3000 | 1331 | 333 | 998 |
| 1985 | 3000 | 1331 | 333 | 998 |
| 1986 | 2500 | 1109 | 277 | 832 |
| 1987 | 2000 | 887 | 222 | 665 |
| 1988 | 2000 | 887 | 222 | 665 |
| 1989 | 2000 | 887 | 222 | 665 |
| 1990 | 1900 | 843 | 211 | 632 |
| 1991 | 1200 | 532 | 133 | 399 |
| 1992 | 2300 | 1020 | 255 | 765 |
| 1993 | 2900 | 1287 | 322 | 965 |
| 1994 | 3400 | 1508 | 377 | 1131 |
| 1995 | 800 | 355 | 89 | 266 |
| 1996 | 1600 | 710 | 177 | 532 |
| 1997 | 1500 | 665 | 166 | 499 |
| 1998 | 2300 | 1020 | 255 | 765 |
| 1999 | 2322 | 1030 | 258 | 773 |
| 2000 | 2267 | 1006 | 251 | 754 |
| 2001 | 2155 | 956 | 239 | 717 |
| 2002 | 1952 | 866 | 217 | 650 |
| 2003 | 2892 | 1283 | 321 | 962 |
| 2004 | 2784 | 1235 | 309 | 926 |
| 2005 | 3287 | 1458 | 365 | 1094 |
| 2006 | 3555 | 1577 | 394 | 1183 |
| 2007 | 1947 | 864 | 216 | 648 |
| 2008 | 3540 | 1571 | 393 | 1178 |
| 2009 | 3460 | 1535 | 384 | 1151 |
| 2010 | 2780 | 1233 | 308 | 925 |
| 2011 | 3757 | 1667 | 417 | 1250 |
| 2012 | 1450 | 643 | 161 | 482 |
| Winbugs labels | SPMHarv[] | Nall_StP\&M | SPMNLarge[] | SPMNSmall[] |

6.iv. Input data for large salmon for Labrador used in the run reconstruction.

| Large salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Commercial harvest |  |  | Proportion Labrador origin |  |  |  |  |  | Exploitation rate |  | Prop. 2SW or 1-fimm |  | Returns to Labrador |  | Angling catches all |  |
| $\begin{aligned} & \text { Year of } \\ & \text { fishery } \end{aligned}$ | SFA 1 | SFA 2 | SFA 14B | SFA 1 |  | SFA 2 |  | SFA 14B |  | All SFAx |  |  |  | Large |  | Large |  |
|  |  |  |  | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Retained | Released |
| *1970 | 1763 | 45479 | 9595 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 562 |  |
| *1971 | 25127 | 64806 | 13673 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 486 | 0 |
| *1972 | 2159 | 55708 | 11753 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 424 |  |
| *1973 | 30204 | 77902 | 16436 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 1009 |  |
| 1974 | 13866 | 93036 | 15863 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 803 |  |
| 1975 | 28601 | 71168 | 14752 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 327 |  |
| 1976 | 38555 | 7796 | 15189 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 830 |  |
| 197 | 28158 | 70158 | 18664 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 1286 |  |
| 1978 | 30824 | 48934 | 11715 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 767 |  |
| 1979 | 21291 | 27073 | 3874 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 609 |  |
| 1980 | 2850 | 87067 | 9138 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 889 |  |
| 1981 | 3614 | 68581 | 7606 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 520 | 0 |
| 1982 | 24192 | 53085 | 5966 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 621 |  |
| 1983 | 19403 | 33320 | 7489 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 428 |  |
| 1984 | 11726 | 25258 | 6218 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 510 | 0 |
| 1985 | 13252 | 16789 | 3954 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 294 |  |
| 1986 | 19152 | 34071 | 5342 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 467 |  |
| 1987 | 1825 | 49799 | 11114 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 633 |  |
| 1988 | 1262 | 32386 | 4591 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 710 | 0 |
| 1989 | 16261 | 26836 | 4646 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 461 |  |
| 1990 | 7313 | 17316 | 2858 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 357 | 0 |
| 1991 | 1369 | 7679 | 4417 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.7 | 0.9 | 0.70 | 0.90 | 0 | 0 | 93 |  |
| 1992 | 9981 | 19608 | 2752 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.5802858 | 0.8295805 | 0.70 | 0.90 | 0 | 0 | 781 | 10 |
| 1993 | 3825 | 9651 | 3620 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.381506 | 0.6232943 | 0.70 | 0.90 | 0 | 0 | 378 | 91 |
| 1994 | 3464 | 11056 | 857 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.2931677 | 0.5039819 | 0.70 | 0.90 | 0 | 0 | 455 | 347 |
| 1995 | 2150 | 8714 | 312 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.1421433 | 0.2513478 | 0.70 | 0.90 | 0 | 0 | 408 | 508 |
| 1996 | 1375 | 5479 | 418 | 0.6 | 0.8 | 0.6 | 0.8 | 0.6 | 0.8 | 0.1269629 | 0.2285846 | 0.70 | 0.90 | 0 | 0 | 334 | 489 |
| *1997 | 1393 | 5550 | 263 | 0.6433 | 0.7247 | 0.8839 | 0.9521 | 0.6 | 0.8 | 0.1700 | 0.3000 | 0.70 | 0.90 | 0 | 0 | 158 | 566 |
| 1998 |  | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 7374 | 19486 | 231 | 814 |
| 1999 | 0 | 0 |  | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 8827 | 23328 | 320 | 931 |
| 2000 | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 12052 | 31850 | 262 | 1446 |
| 2001 | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 12744 | 33677 | 338 | 1468 |
| 2002 | - | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 9076 | 24769 | 207 | 978 |
| 2003 | 0 | 0 | 0 | 1 | 1 | , | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 6676 | 21689 | 222 | 1326 |
| 2004 | 0 | 0 |  | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 10964 | 23092 | 259 | 1519 |
| 2005 | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 11159 | 30796 | 291 | 1290 |
| 2006 | 0 | 0 |  | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 12414 | 29783 | 227 | 1133 |
| 2007 | , | 0 | - | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 11887 | 31913 | 235 | 1222 |
| 2008 | , | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.71 | 1470 | 37677 | 200 | 1461 |
| 2009 | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.70 | 18643 | 60062 | 216 | 1219 |
| 2010 | 0 | 0 | 0 | 1 | 1 | , | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.70 | 7498 | 20099 | 197 | 1080 |
| 2011 | - | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.70 | 8994 | 78695 | 0 | 2114 |
| 2012 | - | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 0.17 | 0.3 | 0.60 | 0.70 | 10054 | 57905 | 0 | 1440 |
| Winbugs variables ] | SFA1- | $\begin{aligned} & \text { LB_SFA2_L } \\ & \text { g_Comm[] } \\ & \hline \end{aligned}$ | $\begin{gathered} \text { LBSSA14B } \\ -\quad \begin{array}{c} \text { Lg_Conmi } \\ 1 \end{array} \end{gathered}$ | $\begin{gathered} \text { pLB_SFA11 } \\ \text { Lg_LI } \end{gathered}$ | $\begin{aligned} & \text { pLB_SFA1_ } \\ & \hline \text { Lg_HD } \end{aligned}$ | $\begin{gathered} \text { pLB_SFA2 } \\ \text { Lg_LI } \end{gathered}$ |  | $\frac{\substack{\text { PLB_SFA } 14 \\ \text { B_Lg_L }}}{(2)}$ | $\begin{gathered} \text { pLB_SFA } 14 \\ \text { B_Lg_HI } \end{gathered}$ | $4 \text { ER-LB-LD_ }$ | $\underset{\mathrm{H}[\mathrm{IC}}{\mathrm{RR}}$ | p2SW_L] | p2SW_H[] | LB_Lg_Lu] | LB_Lg_H[ | $\begin{gathered} \text { LB_Ang } \\ \text { Lg_RetI] } \\ \hline \end{gathered}$ | $\begin{gathered} \text { LB_Ang_ } \\ \text { Lg_Rell }_{4}^{2} \\ \hline \end{gathered}$ |

6.iv. Continued. Input data for small salmon for Labrador used in the run reconstruction.

6.v. Input data for returns of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run reconstruction.

|  | Salmon Fishing Area 3 |  |  |  |  |  | Salmon Fishing Area 4 |  |  |  | Salmon Fishing Area 5 |  |  |  |  | Salmon Fishing Area 6 |  |  |  |  |  | Salmon Fishing Area 7 |  |  |  | Salmon Fishing Area 8 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Small salmon |  |  | Large salmon |  |  | Small salmon |  | Large salmon |  | Small salmon |  | Large salmon |  | $$ |  |  |  | Large salmon |  |  | Small salmon |  | Large salmon |  | Small salmon |  |  | Large salmon |  |
|  | Returns |  | Returns |  |  |  | Returns |  | Returns |  | Returns | Max | Returns |  |  | Returns | Returns | Max |  | Returns Min | Max | Returns Min | Max |  |  |  | Returns | Max |
| Year | Min | Max |  | Min | Max |  | Min | Max | Min | Max | Min |  | Min | Max |  |  |  |  | Min |  |  |  |  |  | Max |  |  |  | Max |  |
| 1970 | 2613 |  | 5227 | 155 |  | 737 | 6163 | 32327 | 957 | 4559 | 7420 | 14840 | 439 |  | 093 |  |  |  |  | 30 | 560 | 17 | 7 | 79 | 67 | 133 |  | 419 | 19 | 62 | 123 |  | $4{ }^{4} 17$ |
| 1971 | 2473 |  | 4947 | 146 |  | 698 | 12610 | 25220 | 746 | 3557 | 5600 | 11200 | 331 |  | 579 | 18 | 13 | 367 | 11 | 1 | 52 | 133 | 267 |  | 8 | 38 | 83 | 167 |  | $5 \quad 24$ |
| 1972 | 1660 |  | 3320 | 98 |  | 468 | 11480 | 22960 | 679 | 3238 | 6317 | 12633 | 374 |  | 1782 | 39 | 97 | 793 |  | 23 | 112 | 203 | 407 | 12 | 12 | 57 | 93 | 187 |  | $6 \quad 26$ |
| 1973 | 3960 |  | 7920 | 234 |  | 1117 | 22367 | 4473 | 1324 | 6308 | 7040 | 14080 | 417 |  | 986 |  | 33 | 1667 |  | 9 | 235 | 437 | 873 | 26 | $26 \quad 123$ | 123 | 313 | 627 | 19 | 98 |
| 1974 | 2797 |  | 5593 | 322 |  | 645 | 17910 | 35820 | 2065 | 4131 | 5457 | 10913 | 629 |  | 1258 | 101 |  | 2020 | 116 | 6 | 233 | 443 | 887 | 51 | 1102 | 102 | 170 | 340 | 20 | 0 |
| 1975 | 3690 |  | 7380 | 520 |  | 1041 | 19810 | 39620 | 2794 | 5587 | 6627 | 13253 | 935 |  | 1869 | 31 |  | 627 | 44 | 4 | 88 | 133 | 267 | 19 | 193 | 38 | 290 | 580 | 41 | 182 |
| 1976 | 3157 |  | 6313 | 380 |  | 760 | 22277 | 4553 | 2683 | 5365 | 6327 | 12653 | 762 |  | 524 |  |  | 1647 |  | 9 | 198 | 100 | 200 | 12 | 12. | 24 | 267 | 533 | 32 | $2 \quad 64$ |
| 1977 | 5100 |  | 10200 | 482 |  | 964 | 27987 | 55973 | 2645 | 5290 | 15387 | 30773 | 1454 |  | 908 | 133 |  | 2673 | 126 |  | 253 | 260 | 520 | ${ }^{25}$ | 5 | 49 | 270 | 540 | 26 | $6 \quad 51$ |
| 1978 | 2527 |  | 5053 | 150 |  | 299 | 2947 | 58493 | 1731 | 3461 | 9527 | 19053 | 564 |  | 128 | 98 |  | 1973 | 58 | 5 | 117 | 330 | 660 | 20 | 0 | 39 | 147 | 293 |  | $9 \quad 17$ |
| 1979 | 6800 |  | 13600 | 390 |  | 779 | 26753 | 53507 | 1533 | 3067 | 4437 | 8873 | 254 |  | 509 | 81 |  | 1627 | 47 | 4 | 93 | 417 | 833 | 24 | 4 | 48 | 333 | 667 | 19 | 9 |
| 1980 | 5810 |  | 11620 | 261 |  | 522 | 31380 | 62760 | 1410 | 2819 | 9007 | 18013 | 405 |  | 809 | 106 |  | 2133 |  | 4 | 96 | 340 | 680 | 15 | 5 | 31 | 400 | 800 | 18 | $8 \quad 36$ |
| 1981 | 7860 |  | 15720 | 1045 |  | 2090 | 45120 | 90240 | 5998 | 11996 | 11627 | 2323 | 1546 |  | 3091 | 201 |  | 4033 | 268 |  | 536 | 410 | 820 | 55 | $5 \quad 109$ | 109 | 257 | 513 | 34 | $4 \quad 68$ |
| 1982 | 8780 |  | 17560 | 212 |  | 424 | 33243 | 66487 | 802 | 1604 | 8110 | 16220 | 196 |  | 391 |  |  | 1920 |  |  | 46 | 517 | 1033 | 12 | 12 | 25 | 283 | 567 |  | $7 \quad 14$ |
| 1983 | 5390 |  | 10780 | 247 |  | 495 | 29847 | 59693 | 1370 | 2740 | 7857 | 15713 | 361 |  | 721 |  |  | 1973 | 45 | 5 | 91 | 463 | 927 | 21 | 1 | 43 | 137 | 273 |  | 6 |
| 1984 | 3532 |  | 7526 | 55 |  | 540 | 34933 | 7436 | 548 | 5337 | 9538 | 20323 | 150 |  | 1457 | 110 |  | 2346 |  |  | 168 | 339 | 722 |  | 5 | 52 | 279 | 594 |  | $4{ }^{43}$ |
| 1985 | 4772 |  | 9879 | 72 |  | 683 | 4408 | 91931 | 671 | 6352 | 12692 | 26275 | 192 |  | 1816 | 156 |  | 3235 |  | 24 | 224 | 408 | 845 |  | 6 | 58 | 375 | 777 |  | $6 \quad 54$ |
| 1986 | 2826 |  | 5898 | 70 |  | 413 | 34015 | -70993 | 840 | 4977 | 14835 | 30963 | 366 |  | 170 | 162 |  | 3400 |  | 4 | 238 | 373 | 779 |  | 9 | 55 | 505 | 1054 | 12 | $12 \quad 74$ |
| 1987 | 2218 |  | 4458 | 57 |  | 318 | 21485 | 43175 | 556 | 3078 | 6556 | 13175 | 170 |  | 939 | 54 |  | 1085 |  | 14 | 77 | 110 | 222 |  | 3 | 16 | 169 | 340 |  | $4 \quad 24$ |
| 1988 | 6624 |  | 13644 | 159 |  | 956 | 37171 | 76566 | 892 | 5367 | 15715 | 32370 | 377 |  | 269 | 161 |  | 3333 |  | 39 | 234 | 483 | 995 | 12 | 12 | 70 | 298 | 614 |  | $7 \quad 43$ |
| 1989 | 3004 |  | 6114 | 90 |  | 461 | 15409 | 31367 | 461 | 2365 | 5767 | 11740 | 172 |  | 885 | 100 |  | 2038 |  | 30 | 154 | 269 | 547 |  | 8 | 41 | 403 | 820 | 12 | 1262 |
| 1990 | 6750 |  | 11816 | 236 |  | 920 | 2224 | 38934 | 776 | 3033 | 9485 | 16602 | 331 |  | 1293 | 131 |  | 2297 |  | 4 | 179 | 193 | 337 |  | 7 | 26 | 338 | 591 | 12 | 1246 |
| 1991 | 5650 |  | 9281 | 193 |  | 750 | 21005 | 34499 | 718 | 2788 | 8793 | 1443 | 301 |  | 167 | 79 |  | 1312 |  | 27 | 106 | 155 | 254 |  | 5 | 21 | 47 | 78 |  | 2 |
| 1992 | 11418 |  | 22836 | 416 |  | 4095 | 38670 | -7739 | 1408 | 13867 | 14189 | 28377 | 516 |  | 088 | 168 |  | 3363 |  | 61 | 603 | 292 | 585 | 11 | $1{ }^{105}$ | 105 | 0 | 0 |  | 0 |
| 1993 | 11793 |  | 22699 | 415 |  | 1614 | 45610 | 87791 | 1605 | 6242 | 16661 | 32071 | 586 |  | 280 | 257 |  | 4954 |  | 1 | 352 | 462 | 890 | 16 | 6 | 63 | 422 | 813 | 15 | $5 \quad 58$ |
| 1994 | 13082 |  | 28738 | 59 |  | 3268 | 29401 | 4585 | 1729 | 734 | - 9740 | 213 | -573 |  | 243 |  |  | 1183 |  |  | 135 | 64 | 141 |  | 4 | 16 | 111 | 243 |  | 7 |
| 1995 | 10205 |  | 24587 | 609 |  | 2665 | 31439 | 75745 | 1877 | 8211 | 1108 | 26762 | 663 |  | 901 |  |  | 931 |  |  | 101 | 233 | 560 |  |  | 61 | 185 | 446 | 11 | $1{ }^{48}$ |
| 1996 | 19519 |  | 43650 | 1439 |  | 4273 | 52515 | 117438 | 3870 | 11497 | 17384 | 38875 | 1281 |  | 8806 |  |  | 1438 |  |  | 141 | 151 | 338 | 11 | 1 | 33 | 224 | 500 | 16 | 6 |
| 1997 | 11763 |  | 21437 | 1226 |  | 3970 | 24074 | 43872 | 2509 | 8125 | 6468 | 11786 | 674 |  | 183 |  |  | 429 |  | 25 | 79 | 60 | 110 |  | 6 | 20 | 60 | 110 |  | $6 \quad 20$ |
| 1998 | 19617 |  | 27571 | 1956 |  | 6992 | 52347 | 73573 | 5219 | 18658 | 11863 | 16673 | 1183 |  | 228 |  |  | 756 |  | 54 | 192 | 249 | 350 | 25 | 5 | 89 | 161 | 227 | 16 | $6 \quad 58$ |
| 1999 | 13981 |  | 20350 | 1286 |  | 4196 | 62141 | 90450 | 5717 | 18651 | 10474 | 15245 | 964 |  | 143 | 40 |  | 589 |  | 37 | 122 | 69 | 100 |  | 6 | 21 | 151 | 220 | 14 | 44 |
| 2000 | 19313 |  | 26033 | 1466 |  | 3728 | 37551 | 50618 | 2850 | 7248 | 12414 | 16734 | 942 |  | 396 | 112 |  | 1520 |  | 8 | 218 | 159 | 214 |  | 12 | 31 | 106 | 143 |  | $8 \quad 20$ |
| 2001 | 11754 |  | 15383 | 907 |  | 2104 | 39901 | 52218 | 3080 | 7143 | 10007 | 13095 | 773 |  | 791 | 29 |  | 387 |  | 23 | 53 | 53 | - 69 |  | 4 | 9 | 20 | 26 |  | 2 |
| 2002 | 10500 |  | 15736 | 684 |  | 2006 | 34310 | 51418 | 2234 | 6556 | 3870 | 5799 | 252 |  | 739 |  |  | 361 |  | 16 | 46 | 0 | 0 |  | 0 | 0 | 72 | 108 |  | $5 \quad 14$ |
| 2003 | 21615 |  | 26166 | 1092 |  | 3485 | 76615 | - 90328 | 3768 | 12032 | 6583 | 7970 | 332 |  | 1062 |  | 58 | 555 |  | 23 | 74 | 104 | - 126 |  | 5 | 17 | 52 | 63 | 3 | 3 |
| 2004 | 7992 |  | 12452 | 396 |  | 1686 | 49598 | -7280 | - 2455 | 10464 | 8385 | 13065 | 415 |  | 1769 | 18 |  | 281 |  | 9 | 38 |  | 0 |  | 0 | 0 | 41 | 64 |  | 2 |
| 2005 | 6421 |  | 1889 | 487 |  | 2678 | 36753 | 108180 | 2790 | 15329 | 5309 | 15627 | 403 |  | 214 | 11 |  | 336 |  | 9 | 48 | 0 | 0 |  | 0 | 0 | 26 | 76 |  | $2 \quad 11$ |
| 2006 | 1075 |  | 17194 | 1251 |  | 3239 | 42745 | -6832 | 4971 | 12872 | 8571 | 13700 | 997 |  | 581 |  | 69 | 110 |  | 8 | 21 | 0 | 0 |  | 0 | 0 | 172 | 275 | 20 | - 52 |
| 2007 | 10422 |  | 21117 | 1182 |  | 3828 | 36934 | 78334 | 4188 | 13567 | 8734 | 17696 | 990 |  | 208 |  | 7 | 157 |  | 9 | 28 | 129 | 262 | 15 | 5 | 47 | 17 | 35 |  | 2 |
| 2008 | 13901 |  | 23285 | 1062 |  | 3396 | 63476 | -106328 | 4851 | 15508 | 11459 | 19195 | 876 |  | 880 |  | 30 | 552 | 25 | 25 | 81 | 84 | - 141 |  | 6 | 21 | 196 | 329 | 15 | 548 |
| 2009 | 13313 |  | 24903 | 787 |  | 5088 | 59555 | 111403 | 3518 | 22760 | 10610 | 19847 | 627 |  | 055 | 48 |  | 908 | 29 | 29 | 185 | 0 | 0 |  | 0 | 0 | 135 | 252 |  | $8 \quad 52$ |
| 2010 | 21058 |  | 26262 | 1610 |  | 4596 | 79694 | 99392 | 6094 | 17393 | 23093 | 28801 | 1766 |  | 5040 | 99 |  | 1243 | 76 | 6 | 218 | 211 | 263 | 16 | 6 | 46 | 110 | 137 | 8 | $8 \quad 24$ |
| 2011 | 15720 |  | 26791 | 1308 |  | 6277 | 60515 | 103137 | 5033 | 24165 | 14418 | 24574 | 1199 |  | 58 | 85 |  | 1448 | 71 | 1 | 339 | 100 | 170 |  | $8 \quad 40$ | 40 | 272 | 464 | 23 | 3109 |
| 2012 | 15959 |  | 26803 | 1355 |  | 5694 | 4 HFAL241 112926 |  | +SFA4Lg_L 5709 SFA4Lg_H |  | 1 14079 | 1 23645 | 1195 |  |  | $3 \quad 737 \quad 123$ |  | $1238 \quad 63$ |  | ${ }_{63}{ }_{\text {LSFAGLg }}$ | ${ }_{\text {H }}{ }^{3} \quad 103$ |  | $3 \quad 174$ | -SFA7LG_L |  | 37HSFA8Sm_LSFA8Sm_H |  |  | $\begin{array}{rrr} 3 & 31 & 130 \\ + \text { SFFA8Lg_L } & \text { SFA8Lg_H } \\ \hline \end{array}$ |  |
| Bugs labe SFA3Sm_L[] S |  | SFA3Sm | _H[] | SFA3LE_LISFA3LI_H |  |  |  |  | Sm_f | SFAGLg_L |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

6.v. Continued. Input data for returns of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run reconstruction.

|  | Salmon Fishing Area 9 |  |  |  | Salmon Fishing Area 10 |  |  |  | Salmon Fishing Area 11 |  |  |  | Salmon Fishing Area 12 |  |  |  | Salmon Fishing Area 13 |  |  |  | Salmon Fishing Area 14A |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | nall salmon |  | Large salmon |  | mall salmon |  | Large salmon |  | mall salmon |  | Large salm |  | mall salmon |  | Large salmon |  | nall salmon |  | Large salmon |  | Small salmon |  | Large salmon |  |
|  | Returns |  | eturns |  | Returns |  | Returns |  | Returns |  | Returns |  | Returns |  | Returns |  | etur |  | Returns |  | eturns |  | etu |  |
| Year | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max |
| 1970 | 6310 | 12620 | 373 | 31780 | 2003 | 4007 | 119 | 565 | 16760 | 33520 | 992 | 4727 | 2497 | 4993 | 148 | - 704 | 25942 | 38282 | 3251 | 5060 | 14817 | 29633 | 365 | 2571 |
| 1971 | 5400 | 10800 | 320 | 1523 | 3093 | 6187 | - 183 | 372 | 13533 | 27067 | 801 | 3817 | 1513 | 3027 | 90 | 427 | 26011 | 40151 | 267 | 475 | 12523 | 25047 | 308 | 2173 |
| 1972 | 3797 | 7593 | 225 | - 1071 | 1890 | 3780 | 112 | 2 533 | 16350 | 32700 | 968 | 4611 | 3093 | 6187 | 183 | - 872 | 23526 | 37589 | 310 | 516 | 8057 | 1611 | 198 | 1398 |
| 1973 | 7200 | 14400 | 426 | - 2031 | 5950 | 11900 | - 352 | 2678 | 16187 | 3237 | 958 | 4565 | 2153 | 4307 | 27 | 607 | 27287 | 40227 | 3303 | 5200 | 17607 | 35213 | ${ }_{433}$ |  |
| 1974 | 4980 | 9960 | 574 | 541149 | 4040 | 8080 | 466 | 932 | 14920 | 29840 | 1720 | 3411 | 2193 | 4387 | 253 | - 506 | 19274 | 28824 | 291 | 425 | 10400 | 20800 | 902 | 1805 |
| 1975 | 6240 | 12480 | 880 | - 1760 | 1423 | 2847 | 201 | 401 | 15003 | 30007 | 2116 | 4232 | 1700 | 3400 | 240 | - 479 | 33671 | 54424 | 497 | 7424 | 16060 | 32120 | 507 | 15 |
| 1976 | 5410 | 10820 | 651 | 1303 | 2433 | 4867 | 293 | 3 566 | 13880 | 27760 | 1671 | 3343 | 990 | 1980 | 119 | - 238 | 29382 | 46902 | 3378 | 5488 | 24603 | 4927 | 1437 | 2874 |
| 1977 | 3600 | 7200 | 340 | 1088 | 3657 | 7313 | 346 | 691 | 13653 | 27307 | 1290 | 2581 | 1860 | 3720 | 176 | - 352 | 17610 | 25240 | 287 | 359 | 19023 | 3847 | 666 | 1331 |
| 1978 | 4343 | 8687 | 257 | 7514 | 5317 | 10633 | 315 | - 629 | 13320 | 26640 | 788 | 1576 | 1220 | 2440 | 72 | 2 144 | 17807 | 27681 | 4716 | 5289 | 10803 | 21607 | 266 | 532 |
| 1979 | 5680 | 11360 | 326 | 6551 | 2830 | 5660 | 162 | - 324 | 11433 | 22867 | 655 | 1311 | 2443 | 4887 | 140 | - 280 | 20372 | 31829 | 1183 | 1862 | 21927 | 43853 | 233 | 467 |
| 1980 | 7930 | 15860 | 356 | 6712 | 5080 | 10160 | 228 | - 456 | 16897 | 33793 | 759 | 1518 | 2733 | 5467 | 123 | 246 | 26538 | 38871 | 5236 | 5913 | 12477 | 24953 | 694 | 1388 |
| 1981 | 6207 | 12413 | 825 | 1650 | 4390 | 8780 | 584 | 41167 | 23540 | 47880 | 3129 | 6258 | 3533 | 7067 | 470 | - 939 | 31359 | 45989 | 5148 | 7452 | 19607 | 39213 | 1090 | 2180 |
| 1982 | 6083 | 12167 | 147 | 293 | 4187 | 8373 | 101 | 202 | 24460 | 48920 | 590 | 1180 | 5183 | 10367 | 125 | 250 | 31628 | 46698 | 3442 | 3831 | 15877 | 31753 | 3094 | 6189 |
| 1983 | 7677 | 15353 | 352 | 2205 | 3800 | 7600 | 174 | 4349 | 15897 | 31793 | 730 | 1460 | 2223 | 4447 | 102 | - 204 | 2082 | 31701 | 4465 | 5100 | 12667 | 25333 | 1704 | 3407 |
| 1984 | 7989 | 1723 | 125 | 1221 | 5141 | 10955 | 81 | 1785 | 24767 | 52774 | 389 | 3784 | 6782 | 14451 | 106 | 61036 | 26184 | 37852 | 2296 | 3710 | 16962 | 36143 | 266 | 2591 |
| 1985 | 375 | 13198 | - 96 | $6 \quad 912$ | 4831 | 10000 | 73 | 3691 | 21213 | 43914 | 320 | 3034 | 3996 | 8273 | 60 | - 572 | 16028 | 25505 | 1375 | 2508 | 13209 | 27345 | 199 | 1890 |
| 1986 | 8411 | 17555 | 208 | $8 \quad 1231$ | 5619 | 11727 | 139 | - 822 | 20300 | 42368 | 501 | 2970 | 3433 | 7166 | 85 | 502 | 22881 | 36916 | 2079 | 3649 | 18411 | 38426 | 455 | 2694 |
| 1987 | 3416 | 6865 | 88 | 8489 | 1690 | 3397 | 4 | $4 \quad 242$ | 15087 | 30317 | 391 | 2162 | 3274 | 6580 | 85 | - 469 | 19629 | 32325 | 1546 | 3022 | 18203 | 36580 | 471 |  |
| 1988 | 5179 | 10668 | 124 | 748 | 4308 | 8873 | 103 | 3622 | 18985 | 39106 | 456 | 2741 | 5330 | 10979 | 128 | 370 | 26162 | 43480 | 1950 | 3917 | 23580 | 48570 | 566 | 3405 |
| 1989 | 5352 | 10895 | 160 | 0821 | 3655 | 7440 | 109 | 961 | 12047 | 24524 | 360 | 1849 | 2279 | 4640 | ${ }_{68}$ | 350 | 10154 | 16156 | 849 | 1565 | 13036 | 26537 | 390 | 2001 |
| 1990 | 7332 | 12834 | 256 | 61000 | 3281 | 5743 | 115 | 447 | 17470 | 30578 | 610 | 2382 | 3363 | 5887 | 117 | 459 | 21518 | 31183 | 1778 | 3084 | 19843 | 3473 | 693 | 2706 |
| 1991 | 2404 | 3949 | -82 | 2319 | 988 | 22 | - 34 | $4 \quad 131$ | 7956 | 13068 | 272 | 1056 | 2765 | 542 | 95 | 567 | 1625 | 20945 | 1709 | 243 | 15307 | 25141 | 523 | 2031 |
| 1992 | 5044 | 10088 | 184 | 4 1809 | 1791 | 3582 | 65 | 5642 | 1615 | 33231 | 605 | 5958 | 4671 | ${ }^{9342}$ | 170 | 1675 | 25990 | 44119 | 3087 | 892 | 34927 | 69854 | 1271 | 225 |
| 1993 | 11402 | 21948 | 401 | 11560 | 5578 | 10736 | 196 | - 763 | 24574 | 47301 | 865 | 3363 | 5936 | 11426 | 209 | -812 | 27523 | 46889 | 2618 | 4746 | 31116 | 59893 | 1095 | 4258 |
| 1994 | 3007 | 6607 | 177 | 751 | 2544 | 5588 | 150 | - 635 | 7649 | 16803 | 450 | 1910 | 2761 | 6066 | 162 | - 690 | 22103 | 37166 | 3476 | 5879 | 13321 | 2926 | 783 | 3327 |
| 1995 | 5321 | 12821 | 318 | $8 \quad 1390$ | 4371 | 10532 | 261 | 1142 | 10757 | 25916 | 642 | 2809 | 2294 | 5527 | 137 | - 599 | 27022 | 49781 | 1843 | 5096 | 20840 | 5029 | 1244 | 5443 |
| 1996 | 6015 | 13450 | 43 | 31317 | 8245 | 18438 | -608 | 1805 | 18938 | 42350 | 1396 | 4146 | 5025 | 11238 | 370 | 1100 | 36576 | 67672 | 3479 | 7132 | 32761 | 73263 | 2415 | - 7172 |
| 1997 | 3636 | 6627 | 379 | 9 1227 | 5071 | 9242 | 528 | - 1712 | 16648 | 30339 | 1735 | 5619 | 4556 | 8303 | 475 | -1538 | 31402 | 46994 | 4240 | 8521 | 25241 | 45998 | 2630 | 9 |
| 1998 | 4694 | 6597 | 468 | 81673 | 7821 | 10992 | 780 | - 2788 | 8467 | 11900 | 844 | 3018 | 2360 | 3318 | 235 | -841 | 21816 | 27955 | 3194 | 7880 | 23995 | 33724 | 2392 | 552 |
| 1999 | 4015 | 5844 | 369 | 91205 | 5113 | 7443 | 470 | - 1535 | 9643 | 14036 | 887 | 2894 | 1139 | 1658 | 105 | -342 | 32407 | 40858 | 3878 | 7739 | 26960 | 39241 | 2480 | 8091 |
| 2000 | 7850 | 10582 | 596 | 61515 | 7639 | 10297 | 580 | - 1475 | 17260 | 23266 | 1310 | 3332 | 2634 | 3551 | 200 | - 509 | 54330 | 67784 | 5519 | 10048 | 36819 | 49632 | 2795 | 7107 |
| 2001 | 2043 | 2674 | 158 | 8366 | 2924 | 3826 | 226 | - 523 | 9396 | 12296 | 725 | 1682 | 2201 | 2880 | 170 | - 394 | 37393 | 45761 | 3749 | 6510 | 20775 | 27188 | 1604 | 3719 |
| 2002 | 1917 | 2873 | 125 | -366 | 3713 | 5565 | 242 | 2 709 | 9011 | 13505 | 587 | 1722 | 2321 | 3478 | 151 | 143 | 34070 | 46011 | 3452 | 6469 | 26558 | 39801 | 1729 | 5075 |
| 2003 | 2229 | 2699 | 113 | $3 \quad 359$ | 3771 | 4565 | 190 | - 608 | 14208 | 17201 | 718 | 2291 | 5917 | 7163 | 299 | - 954 | 50367 | 57997 | 4421 | 8434 | 40802 | 49395 | 2061 | -659 |
| 2004 | 1926 | 3001 | 95 | 5406 | 3697 | 5760 | 183 | 380 | 13762 | 2143 | 681 | 2903 | 3131 | 4879 | 155 | -661 | 4992 | 66549 | 4308 | 9118 | 30057 | 46833 | 1488 | 6341 |
| 2005 | 1948 | 5734 | 148 | $8 \quad 813$ | 2779 | 8180 | 211 | 11159 | 6260 | 18425 | 475 | 2611 | 2686 | 7905 | 204 | + 1120 | 40658 | 88340 | 4595 | 12966 | 17340 | 51040 | 1316 | 7232 |
| 2006 | 4355 | 6960 | 506 | 61311 | 5344 | 8542 | 622 | 1609 | 11033 | 17634 | 1283 | 3322 | 3460 | 5530 | 402 | 1042 | 53311 | 74546 | 8499 | 15058 | 28081 | 4883 | 3266 | 8456 |
| 2007 | 2377 | 4817 | 270 | - 873 | 3497 | 7086 | 397 | - 1285 | 5650 | 11449 | 641 | 2076 | 288 | 5689 | 318 | - 1031 | 33808 | 59140 | 4691 | 10959 | 19966 | 40454 | 2264 | 7334 |
| 2008 | 3944 | 6606 | 301 | $1{ }^{963}$ | 4786 | 8016 | 366 | - 1169 | 11136 | 18654 | 851 | 2721 | 2610 | 4373 | 200 | - 638 | 51933 | 75122 | 3901 | 9668 | 25802 | 4322 | 1972 | 6304 |
| 2009 | 3445 | 6443 | 203 | 31316 | 5137 | 9608 | 303 | 3 1963 | 7536 | 14097 | 445 | 2880 | 1746 | 3266 | 103 | - 667 | 36368 | 55458 | 3722 | 10806 | 21146 | 39555 | 1249 | 8081 |
| 2010 | 6597 | 8227 | 504 | 4 1440 | 8168 | 10187 | 625 | -1783 | 8024 | 10008 | 614 | 1751 | 2999 | 3740 | 229 | -654 | 57930 | 67116 | 5798 | 11067 | 31675 | 39504 | 2422 | 691 |
| 2011 | 5271 | 8983 | 438 | 82105 | 9015 | 15364 | 750 | - 3600 | 6897 | 11755 | 574 | 2754 | 2489 | 4243 | 207 | -994 | 40348 | 68766 | 3356 | 16112 | 24110 | 41092 | 2005 |  |
| 2012 | 3396 | 5703 | 288 | 1212 | 6943 | 11660 | 589 | 2477 | 5924 | 9948 | 503 | 2114 | 2363 | 3968 | 201 | 1843 | 36183 | 60766 | 3072 | 12910 | 27566 | 46295 | 2340 | 9835 |

6.vi. Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run reconstruction.

|  | Salmon Fishing Area 3 |  |  |  | Salmon Fishing Area 4 |  |  |  | Salmon Fishing Area 5 |  |  |  | Salmon Fishing Area 6 |  |  |  | Salmon Fishing Area 7 |  |  |  | Salmon Fishing Area 8 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Small salmon |  | Large salmon |  | Small salmon |  | Large salmon |  | Small salmon |  | Large salmon |  | Small salmon Spawners |  | Large salmon Spawners |  | $\begin{aligned} & \text { Small salmon } \\ & \text { Spawners } \end{aligned}$ |  | $\begin{aligned} & \text { Large salmon } \\ & \text { Spawners } \end{aligned}$ |  | Small salmon Spawners |  | Large salmon Spawners |  |
|  | Spawners |  | Spawners |  | Spawners |  | Spawners |  | Spawne |  | Spawners |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Year | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Spawne Min | Max | Spawners ${ }_{\text {Min }}$ |  |
| 1970 | 1829 | 4443 | 154 | 736 | 11314 | 27478 | 910 | 4512 | 5194 | 12614 | 404 | 2058 | 196 | 476 | 14 | 76 | 47 | 113 |  | $3 \quad 18$ | 43 | 105 | 0 | 13 |
| 1971 | 1731 | 4205 | 135 | 687 | 8827 | 21437 | 688 | 3499 | 3920 | 9520 | 293 | 1541 | 128 | 312 | 10 | 51 | 93 | 227 |  | $8 \quad 38$ | 58 | 142 | 0 | 15 |
| 1972 | 1162 | 2822 | 98 | 468 | 8036 | 19516 | 655 | 3214 | 4422 | 10738 | 354 | 1762 | 278 | 674 | 23 | 112 | 142 | 346 |  | $12 \quad 57$ | 65 | 159 | 6 | 26 |
| 1973 | 2772 | 6732 | 232 | 1115 | 15657 | 38023 | 1275 | 6259 | 4928 | 11968 | 405 | 1974 | 583 | 1417 | 49 | 235 | 306 | 742 | 26 | ${ }^{6} \quad 123$ | 219 | 533 | 15 | 84 |
| 1974 | 1958 | 4754 | 318 | 641 | 12537 | 3047 | 1983 | 4049 | 3820 | 9276 | 608 | 1237 | 707 | 1717 | 115 | 232 | 310 | 754 | 49 | $9 \quad 100$ | 119 | 289 | 20 |  |
| 1975 | 2583 | 6273 | 520 | 1041 | 13867 | 33677 | 2628 | 5421 | 4639 | 11265 | 912 | 1846 | 219 | 533 | 43 | 87 | 93 | 227 |  | $9 \quad 38$ | 203 | 493 | 41 | 82 |
| 1976 | 2210 | 5366 | 379 | 759 | 15594 | 37870 | 2495 | 5177 | 4429 | 1075 | 697 | 1459 | 576 | 1400 | 97 | 196 | 70 | 170 | 12 | $12 \quad 24$ | 187 | 453 | 32 | 64 |
| 1977 | 3570 | 8670 | 478 | 960 | 19591 | 47577 | 1559 | 4204 | 1071 | 26157 | 1410 | 2864 | 936 | 2272 | 107 | 234 | 182 | 442 | 24 | $24 \quad 48$ | 189 | 459 | 26 | 51 |
| 1978 | 1769 | 4295 | 149 | 298 | 20473 | 49719 | 1229 | 2959 | 6669 | 16195 | 536 | 1100 | 691 | 1677 | 51 | 110 | 231 | 561 | 19 | 938 | 103 | 249 | 9 | 17 |
| 1979 | 4760 | 11560 | 390 | 779 | 1872 | 45881 | 1206 | 2740 | 3106 | 7542 | 234 | 489 | 569 | 1383 | 45 | 91 | 292 | 708 | 24 | $4{ }^{48}$ | 233 | 567 | 19 | 38 |
| 1980 | 4067 | 9877 | 224 | 485 | 21966 | 53346 | 903 | 2312 | 6305 | 15311 | 376 | 780 | 747 | 1813 | 34 | 82 | 238 | 578 | 14 | $4{ }^{30}$ | 280 | 680 | 18 | 36 |
| 1981 | 5502 | 13362 | 1042 | 2087 | 31584 | 76704 | 5637 | 11635 | 8139 | 1976 | 1511 | 3056 | 1412 | 3428 | 239 | 507 | 287 | 697 |  | $3 \quad 107$ | 180 | 436 | 34 | ${ }_{6} 8$ |
| 1982 | 6146 | 14926 | 124 | 336 | 23270 | 56514 | 544 | 1346 | 5677 | 13787 | 143 | 338 | 672 | 1632 | 6 | 29 | 362 | 878 |  | $2 \quad 15$ | 198 | 482 | 0 |  |
| 1983 | 3773 | 9163 | 245 | 493 | 2093 | 50739 | 1073 | 2443 | 5500 | 13356 | 191 | 551 | 691 | 1677 | 35 | 81 | 324 | 788 |  | 0 | 96 | 232 | 1 |  |
| 1984 | 2531 | 6525 | 55 | 540 | 25033 | 64536 | 533 | 5322 | 6835 | 17620 | 149 | 1456 | 789 | 2034 | 12 | 163 | 243 | 626 |  | 148 | 200 | 515 | 4 |  |
| 1985 | 3462 | 8569 | 72 | 683 | 32218 | 79741 | 671 | 6352 | 9208 | 22791 | 192 | 1816 | 1134 | 2806 | 24 | 224 | 296 | 733 |  | 6 58 | 272 | 674 | 6 | 54 |
| 1986 | 2054 | 5126 | 70 | 413 | 2472 | 61700 | 840 | 4977 | 10882 | 26910 | 366 | 2170 | 1184 | 2955 | 40 | 238 | 271 | 677 |  | 955 | 367 | 916 | 12 | 74 |
| 1987 | 1655 | 3895 | 57 | 318 | 16032 | 3772 | 556 | 3078 | 4892 | 11511 | 170 | 939 | 403 | 948 | 14 | 77 | 82 | 194 |  | $3 \quad 16$ | 126 | 297 | 4 | 24 |
| 1988 | 4868 | 11888 | 159 | 956 | 27317 | 66712 | 892 | 5367 | 11549 | 28204 | 377 | 2269 | 1189 | 2904 | 39 | 234 | 355 | 867 |  | 1270 | 219 | 535 | 7 |  |
| 1989 | 2266 | 5376 | 90 | 461 | 11623 | 27581 | 461 | 2365 | 4350 | 1032 | 172 | 885 | 755 | 1792 | 30 | 154 | 203 | 481 |  | $8 \quad 41$ | 304 | 721 | 12 |  |
| 1990 | 5032 | 10098 | 236 | 920 | 16583 | 33273 | 776 | 3033 | 7071 | 14188 | 331 | 1293 | 978 | 1963 | 46 | 179 | 144 | 288 |  | $7 \quad 26$ | - 252 | 505 | , | 46 |
| 1991 | 4334 | 7965 | 193 | 750 | 16113 | 29607 | 718 | 2788 | 6745 | 12395 | 301 | 1167 | 613 | 1126 | 27 | 106 | 119 | 218 |  | $5 \quad 21$ | 36 | 67 | 2 |  |
| 1992 | 984 | 21262 | 415 | 4094 | 33228 | 71898 | 1407 | 13866 | 12175 | 26363 | 516 | 5088 | 1450 | 3132 | 61 | 603 | 252 | 545 |  | 1105 | - | 0 | 0 |  |
| 1993 | 10054 | 20961 | 400 | 1599 | 39162 | 8134 | 1590 | 6226 | 14370 | 2977 | 576 | 2270 | 2243 | 4623 | 90 | 351 | 404 | 831 |  | $6 \quad 63$ | 369 | 760 | 15 | 58 |
| 1994 | 9146 | 24802 | 749 | 3247 | 20576 | 55760 | 1644 | 7259 | 6855 | 18510 | 560 | 2420 | 381 | 1026 | 30 | 133 | 46 | 122 |  | $4 \quad 16$ | 69 | 212 | 6 | 27 |
| 1995 | 7409 | 21791 | 580 | 2636 | 22872 | 67179 | 1801 | 8135 | 8122 | 23776 | 642 | 2880 | 287 | 831 | 23 | 100 | 173 | 501 |  | $4{ }^{60}$ | 135 | 397 | 11 | 48 |
| 1996 | 15729 | 39860 | 1412 | 4247 | 42346 | 107268 | 3757 | 11383 | 14095 | 35586 | 1263 | 3787 | 522 | 1317 | 46 | 139 | 124 | 311 |  | $1{ }^{33}$ | 180 | 457 | 16 | 48 |
| 1997 | 9422 | 19095 | 1209 | 3954 | 19309 | 39107 | 2467 | 8083 | 5228 | 10547 | 668 | 2177 | 190 | 384 | 24 | 79 | 49 | 99 |  | $6 \quad 20$ | 48 | 98 | 6 | 20 |
| 1998 | 16390 | 24345 | 1933 | 6969 | 43559 | 64785 | 5160 | 18599 | 9943 | 14753 | 1155 | 4201 | 455 | 673 | 53 | 191 | 212 | 313 | 25 | 588 | 135 | 201 | 16 | 57 |
| 1999 | 11804 | 18173 | 1279 | 4189 | 52390 | 80998 | 5650 | 18583 | 8832 | 13603 | 947 | 3126 | 343 | 528 | 37 | 121 | 58 | 90 |  | $6 \quad 21$ | 119 | 188 | 14 |  |
| 2000 | 17003 | 23723 | 1449 | 3711 | 32879 | 45946 | 2803 | 7201 | 1089 | 15217 | 923 | 2377 | 993 | 1386 | 84 | 217 | 140 | 195 | 12 | 1231 | 88 | 125 | 8 | 20 |
| 2001 | 9861 | 13489 | 892 | 2089 | 33365 | 45682 | 3023 | 7086 | 8344 | 11433 | 767 | 1786 | 250 | 342 | ${ }^{23}$ | 53 | 42 | 59 |  | 4 | 17 | 23 | 2 |  |
| 2002 | 8620 | 13856 | 671 | 1994 | 28999 | 45208 | 2175 | 6498 | 3194 | 5124 | 250 | 737 | 199 | 319 | 15 | 45 | 0 | 0 |  | 0 | $0 \quad 55$ | 91 | 5 | 14 |
| 2003 | 19386 | 23938 | 1085 | 3478 | 67026 | 82739 | 3738 | 12001 | 5926 | 7312 | 331 | 1060 | 412 | 508 | ${ }^{23}$ | 74 | 94 | 116 |  | $5 \quad 17$ | 178 | 58 | 3 |  |
| 2004 | 6942 | 11402 | 390 | 1680 | 43104 | 70785 | 2430 | 10438 | 7307 | 11987 | 412 | 1766 | 158 | 259 | 9 | 38 | 0 | 0 |  | 00 | $0 \quad 35$ | 58 | 2 |  |
| 2005 | 5056 | 17534 | 473 | 2664 | 28996 | 100323 | 2695 | 15235 | 4200 | 14518 | 394 | 2205 | 92 | 314 | 8 | 47 | 0 | 0 |  | 0 | $0 \quad 18$ | 69 | 2 | 11 |
| 2006 | 9402 | 15839 | 1228 | 3216 | 37156 | 62732 | 4925 | 12825 | 7495 | 12623 | 969 | 2554 | 61 | 102 | 8 | 20 | 0 | 0 |  | 0 | $0 \quad 141$ | 244 | 20 | 52 |
| 2007 | 9147 | 19842 | 1171 | 3818 | 3243 | 70143 | 4122 | 13501 | 7641 | 16603 | 978 | 3196 | 68 | 148 | 8 | 28 | 112 | 245 |  | $12 \quad 45$ | 5 15 | 33 | 2 |  |
| 2008 | 11799 | 21183 | 1045 | 3379 | 53591 | 9644 | 4745 | 15402 | 9669 | 17405 | 867 | 2791 | 274 | 497 | 22 | 78 | 69 | 125 |  | $4 \quad 18$ | 159 | 292 | 15 | 48 |
| 2009 | 11205 | 22795 | 779 | 5080 | 49881 | 10172 | 3491 | 22732 | 8828 | 18065 | 622 | 4049 | 412 | 834 | 28 | 185 | , | 0 |  | 00 | 111 | 228 | 7 | 51 |
| 2010 | 18364 | 23569 | 1595 | 4581 | 69075 | 88772 | 6006 | 17304 | 2014 | 25822 | 1754 | 5028 | 874 | 1120 | 76 | 217 | 183 | 235 | 16 | 46 | 93 | 120 | 8 | 24 |
| 2011 | 13193 | 24264 | 1291 | 6261 | 50806 | 93428 | 4789 | 23920 | 12075 | 22230 | 1176 | 5734 | 716 | 1314 | 70 | 339 | 83 | 153 | 8 | 39 | 220 | 412 | 22 | 108 |
| 2012 | 13455 | 24298 | 1324 | 5664 | 56715 | 102401 | 5596 | 23878 | 11844 | 21410 |  |  | $\begin{array}{cc} 618 & 1119 \\ \hline & \text { SFA6SSm } \\ \text { SFA6SSI } \end{array}$ |  | 61 | 261 | $\begin{array}{cc} 86 & 156 \\ \hline \text { ISFA7SSm_SA7SSm } \end{array}$ |  | $\frac{8}{1-S F A 7 S L g \_}$ | $\begin{gathered} 36 \\ \text { _ISFA7SLg_1S S } \\ \hline \end{gathered}$ | $\begin{array}{r} 301 \\ \\ \hline \text { SFA8SS } \end{array}$ | 549 | 31 | 130 |
| nbugs lab, SFA3SSm_L[] |  | SFA3SSm_H | SFA3SLg ISFA3SLg |  | SFA4SSm_SFA4SSm |  | SFAASLE_L |  | SFA5ssm_ |  |  |  | SFA6SLE_SFA6SLE_\| | SFABSSm | FA8SLg | FA8SLE |  |  |  |  |  |  |  |

6.vi (continued). Input data for spawners of small salmon and large salmon for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run reconstruction.

|  | Salmon Fishing Area 9 |  |  |  | Salmon Fishing Area 10 |  |  |  | Salmon Fishing Area 11 |  |  |  | Salmon Fishing Area 12 |  |  |  | Salmon Fishing Area 13 |  |  |  | Salmon Fishing Area 14A |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Small salmon |  | Large salmon |  | Small salmon Spawners |  | Large salmon Spawners |  | Small salmon Spawners |  | Large salmon Spawners |  | Small salmon Spawners |  | Large salmon Spawners |  | Small salmon Spawners |  | Large salmon Spawners |  | Small salmon |  | Large salmon |  |
|  | Spawner |  | Spawners |  |  |  | Spawners |  |  |  | Spawners |  |  |  |  |  |  |  |  |  |
| Year | Min | Max | Min | Max | Min | Max |  |  | Min | Max |  |  | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Min Max |  |
| 1970 | 4417 | 10727 | 361 | 1768 | 1402 | 3406 | 112 | 558 | 11732 | 28492 | 918 | 4653 | 1748 | 424 | 69 | 625 | 1623 | 28543 | 1608 | 3417 | 10372 | 25188 | 134 | 2340 |
| 1971 | 3780 | 9180 | 301 | 1504 | 2165 | 5259 | 166 | 855 | 9473 | 23007 | 736 | 3752 | 1059 | 2573 | 74 | 411 | 16489 | 30629 | 163 | 3705 | 8766 | 21290 | 0 | 185 |
| 1972 | 2658 | 6454 | 217 | 1063 | 1323 | 3213 | 108 | 529 | 11445 | 27795 | 882 | 4525 | 2165 | 5259 | 163 | 852 | 15125 | 2918 | 2004 | 4066 | 5640 | 13696 | 83 | 128 |
| 1973 | 5040 | 12240 | 406 | 2011 | 4165 | 1015 | 310 | 1636 | 11331 | 27517 | 923 | 4530 | 1507 | 3661 | 102 | 582 | 17019 | 29959 | 1911 | 3808 | 1232 | 2993 | 91 | 213 |
| 1974 | 3486 | 8466 | 565 | 1140 | 2828 | 6868 | 452 | 918 | 1044 | 25364 | 1682 | 3403 | 1535 | 3729 | 240 | 493 | 12085 | 21635 | 1997 | 3341 | 7280 | 17680 | 789 | 692 |
| 1975 | 4368 | 10608 | 874 | 1754 | 996 | 2420 | 192 | 392 | 10502 | 25506 | 2076 | 4192 | 1190 | 2890 | 220 | 459 | 21668 | 42421 | 3611 | 6538 | 11242 | 27302 | 417 | 925 |
| 1976 | 3787 | 9197 | 639 | 1291 | 1703 | 4137 | 283 | 576 | 9716 | 23596 | 1629 | 3301 | 693 | 1683 | 114 | 233 | 18999 | 36519 | 2752 | 4862 | 17222 | 41826 | 1337 | 2774 |
| 1977 | 2520 | 6120 | 331 | 671 | 2560 | 6216 | 341 | 686 | 9557 | 23211 | 1272 | 2563 | 1302 | 3162 | 128 | 304 | 10898 | 18528 | 1828 | 2549 | 13316 | 32340 | 194 | 859 |
| 1978 | 3040 | 7384 | 240 | 97 | 3722 | 9038 | 273 | 587 | 9324 | 2264 | 770 | 1558 | 854 | 2074 | 52 | 124 | 12518 | 22392 | 3861 | 4334 | 7562 | 18366 | 194 | 460 |
| 1979 | 3976 | 9656 | 311 | 636 | 1981 | 4811 | 154 | 316 | 8003 | 19437 | 648 | 1304 | 1710 | 4154 | 130 | 270 | 14363 | 25820 | 1070 | 1749 | 15349 | 37275 | 174 | 408 |
| 1980 | 5551 | 13881 | 295 | 651 | 3556 | 8636 | 201 | 429 | 11828 | 28724 | 715 | 1474 | 1913 | 4647 | 94 | 217 | 18625 | 30958 | 4243 | 4920 | 8734 | 21210 | 514 | 1208 |
| 1981 | 4345 | 10551 | 773 | 1598 | 3073 | 7463 | 555 | 1138 | 16478 | 40018 | 3088 | 6217 | 2473 | 6007 | 453 | 922 | 22059 | 36689 | 4485 | 6789 | 13725 | 33331 | 953 | 2043 |
| 1982 | 4258 | 10342 | 114 | 260 | 2931 | 7117 | 91 | 192 | 17122 | 41582 | 537 | 1127 | 3628 | 8812 | 110 | 235 | 22062 | 37132 | 2847 | 3236 | 1114 | 26990 | 2987 | 6082 |
| 1983 | 5374 | 13550 | 281 | 634 | 2660 | 6460 | 95 | 270 | 11128 | 2722 | 703 | 1433 | 1556 | 3780 | 94 | 196 | 1491 | 25364 | 3855 | 4490 | 8867 | 21533 | 1635 | 3338 |
| 1984 | 5725 | 14759 | 120 | 1216 | 3684 | 9498 | 79 | 783 | 17748 | 45755 | 374 | 3769 | 4860 | 12529 | 38 | 968 | 18413 | 30081 | 1987 | 3401 | 12155 | 31336 | 179 | 2504 |
| 1985 | 4625 | 11448 | 96 | 912 | 3505 | 8674 | 73 | 691 | 15390 | 38091 | 320 | 3034 | 2899 | 7176 | 57 | 569 | 10726 | 20203 | 1349 | 2482 | 9583 | 23719 | 197 | 1887 |
| 1986 | 6113 | 15257 | 208 | 1231 | 4084 | 10192 | 139 | 822 | 14754 | 36822 | 501 | 2970 | 2495 | 6228 | 81 | 499 | 15535 | 29570 | 2013 | 3583 | 13381 | 33396 | 445 | 2683 |
| 1987 | 2549 | 5998 | 88 | 489 | 1261 | 2968 | 4 | 242 | 11258 | 26488 | 391 | 2162 | 2443 | 574 | 82 | 466 | 13611 | 26307 | 1512 | 2988 | 13583 | 31960 | 467 | 2604 |
| 1988 | 3806 | 9295 | 124 | 748 | 3166 | 7731 | 103 | 622 | 13952 | 34073 | 456 | 2741 | 3917 | 9566 | 126 | 767 | 17945 | 35263 | 1909 | 3877 | 17329 | 42319 | 549 | 3388 |
| 1989 | 4037 | 9580 | 160 | 821 | 2757 | 6542 | 109 | 561 | 9087 | 21564 | 360 | 1849 | 1719 | 4080 | 67 | 349 | 6980 | 12982 | 836 | 1552 | 983 | 23334 | 385 | 1996 |
| 1990 | 5466 | 10968 | 256 | 1000 | 2446 | 4908 | 115 | 447 | 13024 | 26132 | 610 | 2382 | 2507 | 5031 | 114 | 456 | 14866 | 24531 | 1744 | 3051 | 14793 | 29682 | 679 | 2692 |
| 1991 | 1844 | 3389 | 82 | 319 | 758 | 1392 | 34 | 131 | 6103 | 11215 | 272 | 1056 | 2121 | 3898 | 93 | 365 | 11037 | 15757 | 1689 | 2413 | 11742 | 21576 | 512 | 2020 |
| 1992 | 4334 | 9378 | 183 | 1809 | 1496 | 3287 | 65 | 642 | 14239 | 30854 | 605 | 5958 | 3985 | 8657 | 162 | 1667 | 20506 | 38635 | 2992 | 883 | 30096 | 65023 | 1234 | 12488 |
| 1993 | 9956 | 20502 | 400 | 1559 | 4809 | 9967 | 194 | 761 | 21423 | 4150 | 861 | 3359 | 5176 | 10666 | 207 | 810 | 22341 | 41708 | 2544 | 4673 | 27010 | 55787 | 1058 | 4221 |
| 1994 | 2124 | 5723 | 172 | 746 | 1804 | 4848 | 144 | 630 | 5295 | 1449 | 430 | 1891 | 1949 | 5253 | 154 | 681 | 15381 | 3044 | 3207 | 5611 | 9385 | 25327 | 742 | 3286 |
| 1995 | 3887 | 11386 | 304 | 1376 | 3218 | 9378 | 253 | 1133 | 7770 | 22930 | 625 | 2792 | 1639 | 4922 | 130 | 592 | 20570 | 43329 | 1607 | 4860 | 15218 | 4458 | 1187 | 5385 |
| 1996 | 4868 | 12304 | 431 | 04 | 6687 | 16880 | 592 | 1789 | 15226 | 38638 | 1362 | 4113 | 4082 | 1029 | 358 | 1088 | 29056 | 60152 | 3199 | 6852 | 26584 | 6789 | 2357 | 711 |
| 1997 | 2927 | 5918 | 372 | 1221 | 4086 | 8257 | 519 | 1702 | 13304 | 26995 | 1718 | 5602 | 3655 | 7401 | 464 | 1527 | 25508 | 40599 | 3985 | 8266 | 20359 | 4117 | 2578 | 8467 |
| 1998 | 3937 | 5840 | 458 | 1663 | 6606 | 977 | 771 | 279 | 7024 | 10457 | 836 | 3009 | 1968 | 2925 | 225 | 831 | 18279 | 2417 | 3031 | 6918 | 19992 | 2972 | 2347 | 850 |
| 1999 | 3401 | 5230 | 359 | 1195 | 4313 | 6642 | 455 | 1520 | 8086 | 12478 | 881 | 2889 | 958 | 147 | 102 | 339 | 28647 | 37098 | 3760 | 7621 | 22659 | 34941 | 2402 | 8013 |
| 2000 | 6913 | 9645 | 581 | 1501 | 6664 | 9322 | 534 | 1429 | 14895 | 2091 | 1288 | 3310 | 2291 | 3208 | 195 | 504 | 48055 | 61508 | 5250 | 9779 | 32314 | 45127 | 2731 | 7044 |
| 2001 | 1709 | 2339 | 151 | 359 | 2436 | 3338 | 215 | 513 | 7804 | 10704 | 714 | 1671 | 1818 | 2497 | 162 | 386 | 31037 | 39405 | 3536 | 6297 | 17331 | 2374 | 1559 | 3674 |
| 2002 | 1562 | 2518 | 118 | 360 | 3049 | 4901 | 231 | 699 | 7347 | 11840 | 581 | 1716 | 1896 | 3053 | 147 | 439 | 28083 | 40025 | 3313 | 6330 | 21764 | 35007 | 1668 | 5013 |
| 2003 | 1985 | 2454 | 109 | 355 | 3368 | 4162 | 185 | 603 | 12701 | 15693 | 703 | 2276 | 5282 | 6528 | 288 | 943 | 45027 | 52657 | 4206 | 8218 | 36597 | 45189 | 1988 | 6506 |
| 2004 | 1674 | 274 | 91 | 402 | 3210 | 5273 | 177 | 774 | 11863 | 1954 | 660 | 2882 | 2704 | 442 | 149 | 655 | 43899 | 60513 | 4074 | 8883 | 26116 | 42992 | 1429 | 6282 |
| 2005 | 1478 | 5264 | 130 | 794 | 2171 | 7572 | 194 | 1142 | 4827 | 16992 | 456 | 2591 | 2062 | 7282 | 191 | 1107 | 33349 | 81031 | 4320 | 12691 | 13676 | 47376 | 1246 | 7163 |
| 2006 | 3791 | 6397 | 498 | 1302 | 4627 | 7824 | 602 | 1590 | 9554 | 16155 | 1271 | 3310 | 2986 | 5056 | 392 | 1032 | 46296 | 67532 | 8247 | 14807 | 24532 | 41334 | 3210 | 8400 |
| 2007 | 2063 | 4502 | 263 | 867 | 3047 | 6636 | 387 | 1275 | 4907 | 10706 | 636 | 2071 | 2442 | 5323 | 314 | 1027 | 2942 | 54734 | 4511 | 10780 | 17446 | 37934 | 2222 | 7293 |
| 2008 | 3285 | 5948 | 293 | 955 | 3971 | 7202 | 351 | 1154 | 9314 | 16832 | 841 | 2711 | 2178 | 3940 | 193 | 631 | 43277 | 66465 | 3580 | 9346 | 21887 | 39305 | 1915 | 6246 |
| 2009 | 2835 | 5834 | 198 | 1311 | 4193 | 8665 | 298 | 1957 | 6203 | 12763 | 442 | 2877 | 1450 | 2970 | 100 | 664 | 31106 | 50196 | 3526 | 10610 | 17820 | 36229 | 1200 | 8032 |
| 2010 | 5703 | 7334 | 496 | 1432 | 7062 | 9081 | 616 | 1774 | 6859 | 8842 | 604 | 1742 | 2606 | 3347 | 226 | 651 | 49703 | 58889 | 5478 | 10747 | 27468 | 35298 | 2358 | 6848 |
| 2011 | 4364 | 8077 | 433 | 2099 | 7477 | 13826 | 716 | 3566 | 5696 | 10554 | 564 | 2744 | 2074 | 3827 | 203 | 990 | 33849 | 62267 | 3160 | 15915 | 20249 | 37231 | 1953 | 9575 |
| 2012 | 2824 | 5131 | 280 | 1203 | 5780 | 10497 | 559 | 2447 | 4943 | 8968 | 494 | 2105 | 1990 | 3595 | 198 | 840 | 30562 | 55146 | 2914 | 12752 | 23237 | 41966 | 2283 | 9778 |

6.vii. Input data for 2SW salmon returns and spawners for Salmon Fishing Areas 3 to 8 in Newfoundland used in the run reconstruction.

|  | Salmon Fishing Area 3 |  |  |  | Salmon Fishing Area 4 |  |  |  | Salmon Fishing Area 5 |  |  |  | Salmon Fishing Area 6 |  |  |  | Salmon Fishing Area 7 |  |  |  | Salmon Fishing Area 8 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Returns ${ }^{\text {2SW }}$ | 2Sw | 2sw |  |  |  | $\begin{aligned} & \text { 2su } \\ & \text { Spawners } \\ & \times \quad \text { Min } \end{aligned}$ |  | ${ }_{\text {Returns }}^{2 \mathrm{SW}}$ | $\operatorname{Max}_{\substack{\text { Spawners }}}^{\text {2SW }}$ |  |  | 2sw |  | Spawners <br> Min |  |  | 2SW <br> Spawners |  | Returns |  | ( ${ }_{\text {Spawners }}{ }^{\text {2SW }}$ |  |  |
|  |  |  | wners |  | Returns |  |  |  |  |  |  |  | Returns |  |  |  |  |  |  |  |  |  |  |  |
| Year | Min | Max | Min | Max | Min | Max |  | Max | Min |  |  | Max | Min | Max |  | Max | Min | Max | Min | Max | Min | Max | Min | Max |
| 1970 | 15 | 147 | 15 | 147 | 96 | 912 | 91 | 902 | 44 | 419 | 40 | 412 | 2 | 16 | 1 | 15 | 0 | 4 | 0 | 4 | 0 | 3 | 0 |  |
| 1971 | 15 | 140 | 14 | 137 | 75 | 711 | 69 | 700 | 33 | 316 | 29 | 308 | 1 | 10 | 1 | 10 | 1 | 8 | 1 | 8 | 0 | 5 | 0 |  |
| 1972 | 10 | 94 | 10 | 94 | 68 | 648 | 66 | 643 | 37 | 356 | 35 | 352 | 2 | 22 | 2 | 22 | 1 | 11 | 1 | 11 | 1 | 5 | 1 |  |
| 1973 | 23 | 223 | 23 | 223 | 132 | 1262 | 127 | 1252 | 42 | 397 | 40 | 395 | 5 | 47 | 5 | 47 | 3 | 25 | 3 | 25 | 2 | 18 | 1 | 17 |
| 1974 | 32 | 129 | 32 | 128 | 207 | 826 | 198 | 810 | 63 | 252 | 61 | 247 | 12 | 47 | 12 | 46 | 5 | 20 | 5 | 20 | 2 | 8 | 2 |  |
| 1975 | 52 | 208 | 52 | 208 | 279 | 1117 | 263 | 1084 | 93 | 374 | 91 | 369 | 4 | 18 | 4 | 17 | 2 | 8 | 2 | 8 | 4 | 16 | 4 | 16 |
| 1976 | 38 | 152 | 38 | 152 | 268 | 1073 | 249 | 1035 | 76 | 305 | 70 | 292 | 10 | 40 | 10 | 39 | 1 | 5 | 5 | 5 | 3 | 13 | 3 | 13 |
| 1977 | 48 | 193 | 48 | 192 | 264 | 1058 | 156 | 841 | 145 | 582 | 141 | 573 | 13 | 51 | 11 | 47 | 2 | 10 | 2 | 10 | 3 | 10 | 3 | 10 |
| 1978 | 15 | 60 | 15 | 60 | 173 | 692 | 123 | 592 | 56 | 226 | 54 | 220 | 6 | 23 | 5 | 22 | 2 | 8 | 2 | 8 | 1 | 3 | 1 |  |
| 1979 | 39 | 156 | 39 | 156 | 153 | 613 | 121 | 548 | 25 | 102 | ${ }^{23}$ | 98 | 5 | 19 | 4 | 18 | 2 | 10 | 2 | 10 | 2 | 8 | 2 |  |
| 1980 | 26 | 104 | 22 | 97 | 141 | 564 | 90 | 462 | 40 | 162 | 38 | 156 | 5 | 19 | 3 | 16 | 2 | 6 | 1 | 6 | 2 | 7 | 2 |  |
| 1981 | 104 | 418 | 104 | 417 | 600 | 2399 | 564 | 2327 | 155 | 618 | 151 | 611 | 27 | 107 | 24 | 101 | 5 | 22 | 5 | 21 | 3 | 14 | 3 | 14 |
| 1982 | 21 | 85 | 12 | 67 | 80 | 321 | 54 | 269 | 20 | 78 | 14 | 68 | 2 | 9 | 1 | 6 | 1 | 5 | 50 | 3 | 1 | 3 | 0 |  |
| 1983 | 25 | 99 | 25 | 99 | 137 | 548 | 107 | 489 | 36 | 144 | 19 | 110 | 5 | 18 | 4 | 16 | 2 | 9 | 0 | 2 | 1 | 3 | 0 |  |
| 1984 | 6 | 108 | 6 | 108 | 55 | 1067 | 53 | 1064 | 15 | 291 | 15 | 291 | 2 | 34 | 1 | 33 | 1 | 10 | 0 | 10 | 0 | 9 | 0 |  |
| 1985 | 7 | 137 | 7 | 137 | 67 | 1270 | 67 | 1270 | 19 | 363 | 19 | 363 | 2 | 45 | 2 | 45 | 1 | 12 | 1 | 12 | 1 | 11 | 1 | 11 |
| 1986 | 7 | 83 | 7 | 83 | 84 | 995 | 84 | 995 | 37 | 434 | 37 | 434 | 4 | 48 | 4 | 48 | 1 | 11 | 1 | 11 | 1 | 15 | 1 | 15 |
| 1987 | 6 | 64 | 6 | 64 | 56 | 616 | 56 | 616 | 17 | 188 | 17 | 188 | 1 | 15 | 1 | 15 | 0 | 3 | 0 | 3 | 0 | 5 | 0 |  |
| 1988 | 16 | 191 | 16 | 191 | 89 | 1073 | 89 | 1073 | 38 | 454 | 38 | 454 | 4 | 47 | 4 | 47 | 1 | 14 | 1 | 14 | 1 | 9 | 1 |  |
| 1989 | 9 | 92 | , | 92 | 46 | 473 | 46 | 473 | 17 | 177 | 17 | 177 | 3 | 31 | 3 | 31 | 1 |  | 31 | 8 | 1 | 12 | 1 | 12 |
| 1990 | 24 | 184 | 24 | 184 | 78 | 607 | 78 | 607 | 33 | 259 | 33 | 259 | 5 | 36 | 5 | 36 | 1 | 5 | 51 | 5 | 1 | 9 | 1 |  |
| 1991 | 19 | 150 | 19 | 150 | 72 | 558 | 72 | 558 | 30 | 233 | 30 | 233 | 3 | 21 | 3 | 21 | 1 | 4 | 41 | 4 | 0 | 1 | 0 |  |
| 1992 | 42 | 819 | 42 | 819 | 141 | 2773 | 141 | 2773 | 52 | 1018 | 52 | 1018 | 6 | 121 | 6 | 121 | 1 | 21 | 1 | 21 | 0 | 0 | 0 |  |
| 1993 | 42 | 323 | 40 | 320 | 161 | 1248 | 159 | 1245 | 59 | 456 | 58 | 454 | 9 | 70 | 9 | 70 | 2 | 13 | 32 | 13 | 1 | 12 | 1 | 12 |
| 1994 | 46 | 457 | 45 | 455 | 104 | 1028 | 99 | 1016 | 34 | 341 | 34 | 339 | 2 | 19 | 2 | 19 | 0 | 2 | 20 | 2 | 0 | 4 | 0 |  |
| 1995 | 37 | 373 | 35 | 369 | 113 | 1150 | 108 | 1139 | 40 | 406 | 39 | 403 | 1 | 14 | 1 | 14 | 1 | 9 | 9 | 8 | 1 | 7 | 1 |  |
| 1996 | 86 | 598 | 85 | 595 | 232 | 1610 | 225 | 1594 | 77 | 533 | 76 | 530 | 3 | 20 | 3 | 19 | 1 | 5 | 51 | 5 | 1 | 7 | 1 |  |
| 1997 | 74 | 556 | 73 | 554 | 151 | 1138 | 148 | 1132 | 40 | 306 | 40 | 305 | 1 | 11 | 1 | 11 | 0 | 3 | 30 | 3 | 0 | 3 | 0 |  |
| 1998 | 117 | 979 | 116 | 976 | 313 | 2612 | 310 | 2604 | 71 | 592 | 69 | 588 | 3 | 27 | 3 | 27 | 1 | 12 | 12 | 12 | 1 | 8 | 1 |  |
| 1999 | 77 | 587 | 77 | 586 | 343 | 2611 | 339 | 2602 | 58 | 440 | 57 | 438 | 2 | 17 | 2 | 17 | 0 | 3 | 30 | 3 | 1 | 6 | 1 |  |
| 2000 | 88 | 522 | 87 | 520 | 171 | 1015 | 168 | 1008 | 57 | 335 | 55 | 333 | 5 | 30 | 5 | 30 | 1 | 4 | $4{ }^{1}$ | 4 | 0 | 3 | 0 |  |
| 2001 | 39 | 196 | 38 | 194 | 132 | 664 | 130 | 659 | 33 | 167 | 33 | 166 | 1 | 5 | 1 | 5 | 0 | 1 | 10 | 1 | 0 | 0 | 0 |  |
| 2002 | 29 | 187 | 29 | 185 | 96 | 610 | 94 | 604 | 11 | 69 | 11 | 69 | 1 | 4 | 1 | 4 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |  |
| 2003 | 47 | 324 | 47 | 323 | 162 | 1119 | 161 | 1116 | 14 | 99 | 14 | 99 | 1 | 7 | 1 | 7 | 0 | 2 | 20 | 2 | 0 | 1 | 0 |  |
| 2004 | 17 | 157 | 17 | 156 | 106 | 973 | 104 | 971 | 18 | 165 | 18 | 164 | 0 | 4 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |  |
| 2005 | 21 | 249 | 20 | 248 | 120 | 1426 | 116 | 1417 | 17 | 206 | 17 | 205 | 0 | 4 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |  |
| 2006 | 54 | 301 | 53 | 299 | 214 | 1197 | 212 | 1193 | 43 | 240 | 42 | 237 | 0 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 5 | 1 |  |
| 2007 | 51 | 356 | 50 | 355 | 180 | 1262 | 177 | 1256 | 43 | 298 | 42 | 297 | 0 | 3 | 0 | 3 | 1 | 4 | $4 \quad 1$ | 4 | 0 | 1 | 0 |  |
| 2008 | 46 | 316 | 45 | 314 | 209 | 1442 | 204 | 1432 | 38 | 260 | 37 | 260 | 1 | 7 | 1 | 7 | 0 | 2 | 20 | 2 | 1 | 4 | 1 |  |
| 2009 | 34 | 473 | 33 | 472 | 151 | 2117 | 150 | 2114 | 27 | 377 | 27 | 377 | 1 | 17 | 1 | 17 | 0 | 0 | 0 | 0 | 0 | 5 | 0 |  |
| 2010 | 69 | 427 | 69 | 426 | 262 | 1618 | 258 | 1609 | 76 | 469 | 75 | 468 | 3 | 20 | 3 | 20 | 1 | 4 | 1 | 4 | 0 | 2 | 0 |  |
| 2011 | 56 | 584 | 56 | 582 | 216 | 2247 | 206 | 2225 | 52 | 535 | 51 | 533 | 3 | 32 | 3 | 31 | 0 | 4 | 0 | 4 | 1 | 10 | 1 | 10 |
| 2012 | 58 | 530 | 57 | 527 | 245 | 2231 | 241 | 2221 | 51 | 467 | 51 | 465 | 3 | 24 | 3 | 24 | 0 | 3 | 0 | 3 | 1 | 12 | 1 | 12 |

6.vii. Continued. Input data for 2 SW salmon returns and spawners for Salmon Fishing Areas 9 to 14A in Newfoundland used in the run reconstruction.


## 6.viii. Input data for small salmon returns to Quebec by category of data used in the run reconstruction.


6.viii. Continued. Input data for large salmon returns to Quebec by category of data used in the run reconstruction.

| Large returnsMinimum |  |  |  | C4 | c5 | C6 FN Harvest |  |  |  |  |  |  |  |  | FN Harvest |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | num |  | C3 |  |  |  |  | Other rivers C1 | arge retur |  |  |  |  |  |  |  |
| Year | C1 | c2 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1970 | 0 | 0 | 0 | 0 | , | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | $0$ |  |  |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1972 | 0 | 0 | 0 | 0 | 0 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 14119 | 9501 | 2922 | 3407 | 3712 | 5071 | 1 329 | 108 | 15631 | 9788 | 6035 | 6477 | 6187 | 8452 | 548 | 181 |
| 1985 | 14015 | 7028 | 3836 | 345 | 9215 | 3351 | 329 | 76 | 15611 | 7281 | 7809 | 577 | 15827 | 5586 | 548 | 127 |
| 1986 | 18589 | 8598 | 6152 | 35 | 5877 | 4971 | 1329 | 89 | 20602 | 8839 | 12596 | 61 | 9795 | 8284 | 548 | 149 |
| 1987 | 17574 | 6715 | 5178 | 273 | 6335 | 3012 | 2329 | 82 | 19017 | 6889 | 10575 | 458 | 10558 | 5019 | 548 | 137 |
| 1988 | 21445 | 6432 | 7540 | 346 | 6789 | 4781 | 329 | 98 | 22979 | 6618 | 15336 | 576 | 11315 | 7969 | 548 | 164 |
| 1989 | 20278 | 8503 | 5530 | 278 | 5718 | 4567 | 329 | 106 | 21906 | 8736 | 11252 | 465 | 9531 | 7611 | 548 | 176 |
| 1990 | 17098 | 10803 | 8164 | 1365 | 5179 | 2424 | 442 | 112 | 18222 | 11041 | 16613 | 2276 | 8631 | 4040 | 737 | 187 |
| 1991 | 19112 | 6988 | 7183 | 696 | 3856 | 357 | 242 | 101 | 20443 | 7192 | 14602 | 1161 | 6427 | 595 | 403 | 168 |
| 1992 | 18392 | 7360 | 7930 | 372 | 2687 | 1503 | 461 | 76 | 19578 | 7560 | 16149 | 622 | 4478 | 2505 | 769 | 127 |
| 1993 | 14578 | 10133 | 2866 | 373 | 2649 | 333 | 333 423 | 52 | 15454 | 11463 | 5849 | 624 | 4414 | 555 | 705 | -87 |
| 1994 | 16538 | 9172 | 2644 | 506 | 2853 | 145 | 427 | 60 | 17594 | 10241 | 5411 | 845 | 4755 | 242 | 712 | 1200 |
| 1995 | 21658 | 9598 | 1926 | 813 | 4390 | 154 | 54.246 | 31 | 22968 | 10936 | 3915 | 1358 | 7317 | 256 | 410 | - 52 |
| 1996 | 22679 | 5822 | 3843 | 577 | 2486 | 135 | 135 | 4 | 24117 | 6941 | 7844 | 964 | 4155 | 225 | 189 |  |
| 1997 | 18106 | 4221 | 2816 | 333 | 2865 | 138 | - 48 | 3 | 19154 | 5154 | 5768 | 553 | 4775 | 229 | 80 | - 15 |
| 1998 | 13180 | 4927 | 2861 | 347 | 2790 | 291 | - 48 | 30 | 13891 | 5962 | 5907 | 592 | 4649 | 485 | 80 |  |
| 1999 | 16912 | 842 | 2554 | 3661 | 3870 | 492 | 20 | 0 | 17700 | 995 | 5232 | 6103 | 6450 | 838 | 0 | 0 |
| 2000 | 14568 | 619 | 3901 | 560 | 6420 | 563 | 530 | 0 | 15300 | 669 | 7947 | 933 | 10700 | 949 | 0 | 0 |
| 2001 | 17837 | 633 | 5320 | 241 | 3988 | 556 | 0 | 0 | 18889 | 879 | 10914 | 402 | 6647 | 926 | 0 | 0 |
| 2002 | 12335 | 8 | 4515 | 339 | 2103 | 345 | 45 0 | 0 | 13001 | 9 | 9277 | 565 | 3505 | 575 | 0 | 0 |
| 2003 | 21853 | 0 | 5787 | 269 | 4889 | 384 | 340 | 0 | 22893 | 0 | 11779 | 449 | 8148 | 641 | 0 | 0 |
| 2004 | 18369 | 107 | 4870 | 357 | 4432 | 401 | 10 | 0 | 19043 | 126 | 9170 | 595 | 7387 | 668 | 0 | 0 |
| 2005 | 19154 | 0 | 3204 | 734 | 4815 | 351 | 510 | 0 | 20066 | 0 | 6515 | 1223 | 8025 | 585 | 0 | 0 |
| 2006 | 16704 | 0 | 3387 | 901 | 3945 | 403 | 03 | , | 17500 | 0 | 6904 | 1502 | 6575 | 672 | 0 | 0 |
| 2007 | 14832 | 0 | 3638 | 1301 | 3171 | 305 | 05 | 0 | 15604 | 0 | 7406 | 2168 | 5285 | 508 | 0 | 0 |
| 2008 | 15216 | 0 | 5187 | 1328 | 5423 | 390 | 0 | 0 | 16002 | 0 | 10595 | 2213 | 9038 | 649 | 0 | 0 |
| 2009 | 18479 | 0 | 3727 | 950 | 4556 | 275 | 2750 | 0 | 19412 | 0 | 7589 | 1584 | 7594 | 458 | 0 | 0 |
| 2010 | 21375 | 0 | 4488 | 1047 | 3656 | 338 | - | 0 | 22454 | 0 | 9157 | 1744 | 6093 | 564 | 0 | 0 |
| 2011 | 26977 | 0 | 4697 | 1571 | 5574 | 483 | 33 | 0 | 28373 | 0 | 9529 | 2619 | 9290 | 805 | 0 | 0 |
| 2012 | 17917 | 0 | 3665 | 787 | 4490 | 367 | 57 0 | 0 | 18836 | 0 | 7434 | 1311 | 7483 | 612 | 0 |  |
| abels | C1_L[] | C2_L] | [3-L[] | C4_L] | C5_L] | C6_L] | QCLgFn_LI | QCLgO_L[] | gC1_H[] ${ }^{\text {a }}$ | C2_H[] | C3_H[] | C4_H[] | CC5_H[] | C6_H[] | QCLgFn_H[] | QCLIgO_H[] |

6.viii. Continued. Input data for small salmon spawners to Quebec by category of data used in the run reconstruction.

6.viii. Continued. Input data for large salmon spawners to Quebec by category of data used in the run reconstruction.

6.viii. Continued. Year specific harvest data (1984 to 2009) and returns and spawners data for Quebec for years when category splits are not available (1970 to 1983 ) used in the run reconstruction.

|  |  |  |  |  |  | These data are specific to the 1970 to 1983 period when detailed returns by river category are not available. |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Harvests in various fishermall salmon |  | ries not in the other inputsLarge salmon |  |  |  | Small returns |  | Large returns |  | Small spawners |  | Large spawners |  |
| Year | Sport | FN | Commercial | Sport | fN | Commercial | Min | Max | Min | Max | Min | Max | Min | Max |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 | 18904 | 28356 | 82680 | 124020 | 11045 | 16568 | 31292 | 46937 |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 14969 | 22453 | 47354 | 71031 | 9338 | 14007 | 16194 | 24292 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 12470 | 18704 | 61773 | 92660 | 8213 | 12320 | 31727 | 47590 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 16585 | 24877 | 68171 | 102256 | 10987 | 16480 | 32279 | 48419 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 16791 | 25186 | 91455 | 137182 | 10067 | 15100 | 39256 | 58884 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 18071 | 27106 | 77664 | 116497 | 11606 | 17409 | 32627 | 48940 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 19959 | 29938 | 77212 | 115818 | 12979 | 19469 | 31032 | 46548 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 18190 | 27285 | 91017 | 136525 | 12004 | 18006 | 44660 | 66990 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 16971 | 25456 | 81953 | 122930 | 11447 | 17170 | 40944 | 61416 |
| 1979 | 0 | 0 | 0 | 0 | 0 | , | 21683 | 32524 | 45197 | 67996 | 15863 | 23795 | 17543 | 26315 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 29791 | 44686 | 107461 | 16192 | 20817 | 31226 | 48758 | 73137 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 41667 | 62501 | 84428 | 126642 | 30952 | 46428 | 35798 | 53697 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 23699 | 35549 | 74870 | 112305 | 16877 | 25316 | 36290 | 54435 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 17987 | 26981 | 61488 | 92332 | 12030 | 18045 | 23710 | 35565 |
| 1984 | 3492 | 357 | 794 | 8561 | 4530 | 13053 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1985 | 4046 | 273 | 2093 | 9883 | 3623 | 16619 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1986 | 6266 | 372 | 3707 | 11643 | 4519 | 20889 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 7443 | 366 | 2992 | 9740 | 4466 | 22745 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1988 | 8663 | 397 | 4760 | 12980 | 4747 | 19750 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1989 | 6080 | 196 | 2615 | 11040 | 2905 | 18175 | 0 | , | 0 | 0 | 0 | 0 | 0 |  |
| 1990 | 8581 | 108 | 3425 | 12132 | 2900 | 16092 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1991 | 6271 | 265 | 3282 | 11194 | 4335 | 16372 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1992 | 8263 | 120 | 3849 | 12291 | 4550 | 15851 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1993 | 8319 | 7 | 3627 | 9798 | 3976 | 11242 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1994 | 7655 | 161 | 3861 | 10932 | 4496 | 10424 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 4187 | 353 | 3915 | 7892 | 6194 | 10038 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1996 | 7265 | 72 | 4532 | 9618 | 6113 | 7454 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1997 | 5075 | 35 | 3531 | 6771 | 4875 | 7202 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 5867 | 35 | 1068 | 4702 | 4875 | 1038 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1999 | 4428 | 710 | 814 | 4407 | 3683 | 471 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 2000 | 5553 | 821 | 0 | 4297 | 3818 | - | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 4213 | 770 | 0 | 5558 | 3574 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 7206 | 1672 | 0 | 2484 | 3164 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 4898 | 972 | 0 | 4610 | 3541 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 6633 | 1158 | 0 | 4412 | 3558 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 3767 | 909 | 0 | 3973 | 3062 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 5366 | 1117 | 0 | 3032 | 3512 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 3787 | 869 | 0 | 3419 | 2932 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 7604 | 1171 | 0 | 3038 | 2971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 3444 | 1141 | 0 | 3338 | 2752 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 4917 | 1057 | 0 | 3166 | 2362 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 7298 | 1205 | 0 | 4295 | 3216 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 4044 | 1224 | 0 | 2740 | 2963 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

6.ix. Input data for 2SW salmon returns to Salmon Fishing Areas $\mathbf{1 5}$ to $\mathbf{2 3}$ for Canada and for USA used in the run reconstruction.

6.ix. Continued. Input data for large salmon returns to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run reconstruction.

| Returns of large salmon |  |  |  | SFA 17 |  | SFA 18 |  |  |  | SFA 19-21 |  | SFA 23 |  | USA |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SFA 15 | Max | SFA 16 |  |  |  |  |  |  |  |  |
| $\begin{gathered} \text { Year of return } \\ \text { to rivers } \end{gathered}$ | Min |  | Min | Max | Min |  |  |  | Max | Min |  | Max | Min | Max | Min | Max | Point estimate |
| 1970 | 12681 | 16270 | 46462 | 49599 | 31 |  | 60 | 6161 | 7858 | 7273 | 9671 | 9691 | 13945 |  |
| 1971 | 5518 | 7102 | 28365 | 33409 | 29 |  | 29 | 2456 | 3198 | 5350 | 6773 | 8056 | 11573 | 653 |
| 1972 | 8441 | 16536 | 30146 | 45087 | 402 |  | 402 | 6095 | 6924 | 7460 | 9082 | 8890 | 12536 | 383 |
| 1973 | 8393 | 16229 | 27771 | 42276 | 206 |  | 206 | 5376 | 6299 | 8049 | 10069 | 4760 | 6638 | 1427 |
| 1974 | 9950 | 19959 | 43249 | 66179 | 386 |  | 386 | 7119 | 7963 | 13138 | 15363 | 12187 | 16444 | 1394 |
| 1975 | 5510 | 10028 | 29826 | 45305 | 345 |  | 345 | 4483 | 498 | 12261 | 13797 | 14829 | 20351 | 23 |
| 1976 | 9596 | 18969 | 23943 | 36016 | 575 |  | 578 | 3578 | 4223 | 8873 | 10416 | 16128 | 22175 | 1317 |
| 1977 | 11053 | 21779 | 52673 | 77434 | 606 |  | 606 | 5175 | 6280 | 14119 | 16690 | 19165 | 26183 | 1998 |
| 1978 | 7277 | 14332 | 22653 | 30245 | 0 |  | 0 | 5954 | 7201 | 10471 | 12378 | 9335 | 12342 | 4208 |
| 1979 | 2886 | 4971 | 9435 | 13507 | 459 |  | 463 | 1676 | 2315 | 5180 | 6684 | 5856 | 7903 | 1942 |
| 1980 | 8768 | 17340 | 37014 | 51008 | - 2 |  | 5 | 4846 | 5951 | 16388 | 20137 | 21464 | 29480 | 5796 |
| 1981 | 9729 | 15652 | 16708 | 28887 | 40 |  | 77 | 3234 | 4332 | 11706 | 15501 | 12481 | 16743 | 5601 |
| 1982 | 7311 | 10700 | 26504 | 51475 | 16 |  | 31 | 5370 | 6783 | 9485 | 11390 | 11147 | 15303 | 6056 |
| 1983 | 5852 | 8950 | 20309 | 35304 | 17 |  | 32 | 4848 | 6024 | 6562 | 8501 | 10908 | 15235 | 215 |
| 1984 | 4214 | 7711 | 12941 | 33321 | 13 |  | 26 | 3105 | 4107 | 2408 | 3909 | 17706 | 24992 | 3222 |
| 1985 | 7627 | 15080 | 16798 | 43247 | - 8 |  | 15 | 1196 | 5150 | 8512 | 14968 | 18582 | 26289 | 5529 |
| 1986 | 10305 | 20267 | 25342 | 65228 | - 5 |  | 11 | 2953 | 13195 | 10722 | 18854 | 11142 | 15761 | 6176 |
| 1987 | 7556 | 14255 | 15734 | 40483 | 66 |  | 128 | 3391 | 11731 | 5950 | 10462 | 7865 | 11116 | 3081 |
| 1988 | 9933 | 19441 | 17627 | 45267 | 96 |  | 185 | 3289 | 11486 | 7321 | 12891 | 5360 | 7312 | 328 |
| 1989 | 7701 | 14898 | 13955 | 35812 | 149 |  | 287 | 2738 | 9680 | 6969 | 12275 | 7393 | 10380 | 3197 |
| 1990 | 6362 | 12307 | 23164 | 59479 | 284 |  | 545 | 2458 | 8649 | 6191 | 10897 | 6235 | 8710 | 5051 |
| 1991 | 4773 | 9335 | 24273 | 62373 | 188 |  | 361 | 3052 | 10837 | 4112 | 7240 | 8312 | 11659 | 2647 |
| 1992 | 7411 | 14420 | 34573 | 49886 | 95 |  | 183 | 3083 | 10780 | 3657 | 6437 | 7749 | 10726 | 2459 |
| 1993 | 3487 | 6711 | 22602 | 87407 | 22 |  | 43 | 1742 | 5985 | 3218 | 5658 | 5260 | 5980 | 2231 |
| 19 | 66 | 12908 | 18098 | 2992 | 269 |  | 310 | 2573 | 9110 | 1743 | 3071 | 3659 | 4155 | 1346 |
| 1995 | 4171 | 8199 | 30324 | 44094 | 384 |  | 576 | 1946 | 6934 | 2532 | 4460 | 3728 | 4289 | 1788 |
| 1996 | 6026 | 11929 | 16317 | 28035 | 394 |  | 591 | 4217 | 15204 | 3571 | 6283 | 5535 | 6365 | 2407 |
| 1997 | 3828 | 7535 | 14711 | 24521 | 387 |  | 581 | 4443 | 16044 | 1550 | 2726 | 3210 | 3678 | 1611 |
| 1998 | 2595 | 5015 | 14774 | 26094 | 385 |  | 577 | 2669 | 9634 | 1359 | 1867 | 2032 | 2437 | 1526 |
| 1999 | 2738 | 5269 | 14550 | 22855 | 383 |  | 575 | 2022 | 7424 | 1709 | 2350 | 2734 | 3090 | 1168 |
| 2000 | 3493 | 6785 | 15734 | 25199 | 378 |  | 566 | 1905 | 7093 | 1315 | 1809 | 1189 | 1430 | 533 |
| 2001 | 5815 | 11449 | 22423 | 30330 | 376 |  | 564 | 2194 | 8162 | 1980 | 2724 | 2113 | 2501 | 797 |
| 2002 | 3592 | 6985 | 10980 | 18385 | 372 |  | 557 | 1595 | 5974 | 749 | 1029 | 639 | 752 | 526 |
| 2003 | 6072 | 11966 | 18726 | 29192 | 371 |  | 557 | 3091 | 11489 | 1952 | 2682 | 1128 | 1289 | 1199 |
| 2004 | 4623 | 9055 | 18669 | 32446 | 367 |  | 550 | 3427 | 12828 | 1302 | 1789 | 1402 | 1698 | 1316 |
| 2005 | 5265 | 10346 | 16769 | 30808 | 373 |  | 560 | 2879 | 10461 | 860 | 1177 | 890 | 1121 | 99 |
| 2006 | 3924 | 7651 | 18680 | 32412 | 392 |  | 587 | 2746 | 10142 | 1559 | 2141 | 997 | 1276 | 1030 |
| 2007 | 6565 | 12957 | 16007 | 26166 | 412 |  | 618 | 1900 | 7062 | 701 | 959 | 689 | 841 | 958 |
| 2008 | 4382 | 8572 | 10427 | 20023 | 429 |  | 644 | 2843 | 10834 | 1928 | 2650 | 858 | 1105 | 1799 |
| 2009 | 6074 | 11970 | 16985 | 27486 | 402 |  | 602 | 1789 | 7344 | 1034 | 1418 | 1678 | 2158 | 2095 |
| 2010 | 4581 | 8972 | 15848 | 26394 | 439 |  | 658 | 2793 | 10825 | 1061 | 1451 | 1117 | 1398 | 1098 |
| 2011 | 11177 | 22223 | 25301 | 58840 | 653 |  | 980 | 4869 | 19502 | 1504 | 2065 | 2598 | 3421 | 3087 |
| 2012 | 4969 | 9750 | 11299 | 22105 | 415 |  | 622 |  | 4344 | 786 | 1073 |  |  |  |
| Winbugs labels | $\stackrel{\text { SF15Lg_L }}{\mathrm{D}}$ | $\begin{array}{r} \mathrm{SF} 15 \mathrm{Lg} \\ \mathrm{HI} \\ \hline \end{array}$ | $\mathrm{a}^{\text {SF16Lg_L }}$ | $\underset{\mathrm{H}[\mathrm{I}}{\mathrm{SF} 16 \mathrm{Lg}}$ | $\mathbf{n}^{\text {SF17Lg_L }}$ | $\begin{array}{r} \text { SF17Lg } \\ \hline \end{array}$ | $\begin{aligned} & 7 \mathrm{Lg} \text { SF } \\ & \begin{array}{c} \mathrm{HID} \end{array} \end{aligned}$ | F18Lg-L | $\begin{array}{\|} \mathrm{SF}^{\mathrm{S} 1 \mathrm{Lg}} \mathrm{I}[\mathrm{~S} \\ \hline \end{array}$ | $\begin{array}{r} \mathrm{SF} 19 \_214 \\ \mathbf{g}_{1} \mathrm{Lu} \\ \hline \end{array}$ |  | SF23LG_L | $\begin{array}{r} \mathrm{SF} 23 \mathrm{Lg} \\ \mathrm{HID} \end{array}$ | USALg] |

6.ix. Continued. Input data for small salmon returns to Salmon Fishing Areas $\mathbf{1 5}$ to $\mathbf{2 3}$ for Canada and for USA used in the run reconstruction.

6.ix. Continued. Input data for 2SW salmon spawners to Salmon Fishing Areas $\mathbf{1 5}$ to $\mathbf{2 3}$ for Canada and for USA used in the run reconstruction.

| Year of returnto rivers | Spawners of 2SW |  | SFA 16 |  | SFA 17 | SFA 18 |  |  |  | SFA 19-21 |  | SFA 23 |  | USA |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SFA 15 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Min | Max | Min | Max | Min | Max | Min |  | Max | Min | Max |  | Min | Max | Point estimate |
| 1970 | 1156 | 3252 | 5346 | 8242 | 18 |  | 47 | 304 | 1587 | 2388 | 4234 | 1536 | 4846 |  |
| 1971 | 510 | 1434 | 6724 | 11354 | 0 |  | 0 | 133 | 694 | 1418 | 2513 | 3612 | 6576 | 490 |
| 1972 | 2367 | 6656 | 2031 | 31450 | 0 |  | 0 | 148 | 775 | 1616 | 2865 | 6472 | 9806 | 1038 |
| 1973 | 2873 | 8081 | 19277 | 33170 | 0 |  | 0 | 165 | 863 | 2246 | 3984 | 2752 | 4412 | 100 |
| 1974 | 3620 | 10183 | 31192 | 52012 | - |  | 0 | 151 | 790 | 2878 | 5103 | 8123 | 12046 | 147 |
| 1975 | 1769 | 4975 | 18536 | 31972 | 0 |  | 0 | 91 | 473 | 1987 | 3523 | 10987 | 16209 | 1942 |
| 1976 | 3530 | 9928 | 11842 | 22152 | 1 |  | 4 | 116 | 604 | 1935 | 343 | 10071 | 15583 | 26 |
| 1977 | 4412 | 12408 | 30623 | 54071 | - 0 |  | 0 | 198 | 1033 | 2559 | 4539 | 12013 | 18568 | 643 |
| 1978 | 2622 | 7375 | 6998 | 13535 | - |  | 0 | 223 | 1166 | 1948 | 3455 | 5346 | 8076 | 314 |
| 1979 | 527 | 1482 | 3000 | 5806 | 3 |  | 7 | 115 | 598 | 1419 | 2517 | 3772 | 5650 | 1509 |
| 1980 | 3440 | 9677 | 17667 | 30961 | 1 |  | 4 | 198 | 1033 | 4170 | 7394 | 12023 | 19005 | 4263 |
| 1981 | 1380 | 3880 | 2392 | 10515 | 36 |  | 73 | 196 | 1027 | 3631 | 6439 | 3642 | 7014 | 4334 |
| 1982 | 991 | 2786 | 8418 | 28619 | -8 |  | 23 | 253 | 1322 | 1158 | 2053 | 4475 | 7939 | 4643 |
| 1983 | 906 | 2547 | 5516 | 17586 | 15 |  | 30 | 210 | 1100 | 1579 | 2800 | 468 | 3561 | 1769 |
| 1984 | 2656 | 5402 | 11650 | 30889 | 13 |  | 26 | 259 | 1148 | 1416 | 2512 | 12280 | 18798 | 2547 |
| 1985 | 4514 | 9180 | 14019 | 37030 | - 8 |  | 15 | 871 | 4359 | 6761 | 11990 | 11885 | 18624 | 4884 |
| 1986 | 7279 | 14804 | 20606 | 54630 | - 5 |  | 11 | 2164 | 11213 | 6624 | 11748 | 7224 | 11280 | 5570 |
| 1987 | 4122 | 8383 | 11414 | 31114 | 66 |  |  | 2534 | 9977 | 3676 | 6519 | 5628 | 8597 | 2781 |
| 1988 | 6582 | 13386 | 13801 | 36355 | 96 |  |  | 2451 | 9748 | 4322 | 7664 | 3420 | 5248 | 3038 |
| 1989 | 3944 | 8021 | 3466 | 22739 | 149 |  |  | 2042 | 8222 | 4735 | 8396 | 6310 | 9158 | 2800 |
| 1990 | 3886 | 7903 | 13669 | 36039 | 284 |  |  | 1829 | 7336 | 3530 | 6260 | 4926 | 7292 | 4356 |
| 1991 | 2193 | 4460 | 14200 | 37251 | 188 |  |  | 2275 | 9204 | 2912 | 5165 | 6080 | 9158 | 241 |
| 1992 | 3639 | 7400 | 20770 | 30116 | 95 |  |  | 2291 | 9131 | 2588 | 4589 | 5826 | 8633 | 229 |
| 1993 | 1239 | 2521 | 15239 | 59907 | 22 |  | 43 | 1296 | 5072 | 2493 | 4421 | 3291 | 3654 | 2065 |
| 1994 | 3639 | 7401 | 13418 | 24653 | 166 |  |  | 1920 | 7743 | 1339 | 2375 | 2387 | 2680 | 1344 |
| 1995 | 2519 | 5124 | 25326 | 36949 | 380 |  |  | 1453 | 897 | 218 | 3934 | 26 | 3652 | 1748 |
| 1996 | 3688 | 7502 | 10743 | 18662 | 388 |  |  | 3166 | 12987 | 2946 | 5224 | 4009 | 4585 | 2407 |
| 1997 | 2316 | 4710 | 8106 | 13754 | 385 |  |  | 3334 | 13698 | 1140 | 2022 | 2219 | 2565 | 1611 |
| 1998 | 1512 | 3076 | 5921 | 10562 | 382 |  |  | 2000 | 8216 | 915 | 1261 | 1068 | 1302 | 1526 |
| 1999 | 1581 | 3217 | 6572 | 10578 | 379 |  |  | 1523 | 6359 | 1409 | 1941 | 1934 | 2181 | 1168 |
| 2000 | 2057 | 4184 | 7160 | 11606 | 376 |  | 56 | 1438 | 6085 | 1072 | 1477 | 805 | 1004 | 1587 |
| 2001 | 3521 | 7161 | 13906 | 18965 | 374 |  |  | 1654 | 6995 | 1812 | 2497 | 1699 | 2008 | 1491 |
| 2002 | 2120 | 4312 | 5275 | 8961 | 371 |  |  | 1203 | 5121 | 378 | 521 | 317 | 356 | 511 |
| 2003 | 3683 | 7491 | 10560 | 16629 | 368 |  |  | 2333 | 9854 | 1834 | 2528 | 878 | 998 | 1192 |
| 2004 | 2770 | 5633 | 10189 | 17971 | 365 |  |  | 2581 | 10986 | 1017 | 1401 | 1238 | 1492 | 1283 |
| 2005 | 3175 | 6457 | 10597 | 19849 | 371 |  | 50 | 2162 | 8935 | 646 | 890 | 726 | 914 | 1088 |
| 2006 | 2329 | 4737 | 9271 | 16348 | 390 |  |  | 2062 | 8667 | 1248 | 1720 | 796 | 1023 | 1419 |
| 2007 | 3994 | 8124 | 8956 | 14896 | 409 |  |  | 1431 | 6047 | 587 | 809 | 530 | 633 | 1189 |
| 2008 | 2618 | 5325 | 5404 | 10730 | 429 |  |  | 2131 | 9252 | 1778 | 2450 | 736 | 953 | 2809 |
| 2009 | 3684 | 7494 | 10013 | 16465 | 401 |  | 02 | 1335 | 6263 | 811 | 1118 | 1391 | 1774 | 2292 |
| 2010 | 2743 | 5580 | 7611 | 12941 | 438 |  | 58 | 2100 | 9266 | 910 | 1253 | 726 | 877 | 1482 |
| 2011 | 6902 | 14038 | 20732 | 48983 | 652 |  |  | 3659 | 16698 | 1467 | 2023 | 2430 | 3196 | 3872 |
| 2012 | 2988 | 6077 | 7525 | 15021 |  |  |  |  |  |  | 883 |  | 298 |  |
| Winbugs labels | $\begin{array}{\|c} \text { SF15S2 } \\ \text { L } \\ \hline \end{array}$ | $\begin{gathered} \mathrm{SF}_{15 \mathrm{~S}}^{2} \\ \mathrm{HI} \\ \hline \end{gathered}$ | $\begin{array}{r} \text { SF16S2 } \\ \text { LI } \end{array}$ | ${ }_{\text {SF16S2 }}^{\text {Hil }}$ | $\begin{gathered} \text { SF172 } \\ \hline 1 \end{gathered}$ | SF1752 | $2^{20} 5$ | $\begin{aligned} & \text { SF18S2 } \\ & \text { LI } \end{aligned}$ | $\begin{array}{r} \mathrm{SF} 18 \mathrm{~S} 2 \\ \mathrm{H}[\mathrm{I} \end{array}$ | $\begin{array}{r} \text { SF19-21 } \\ \text { S2 } 21 \end{array}$ | $\begin{gathered} \mathrm{SF} 1921 \\ \mathrm{~S} 2 \mathrm{HI} \end{gathered}$ | $\begin{array}{\|c} \mathrm{SF} 23 \mathrm{~S} 2 \\ \mathrm{LD} \end{array}$ | $\underset{\mathrm{H}[\mathrm{I}}{\mathrm{SF} 23 \mathrm{~S} 2}$ | USAS20 |

6.ix. Continued. Input data for large salmon spawners to Salmon Fishing Areas 15 to 23 for Canada and for USA used in the run reconstruction.

6.ix. Continued. Input data for small salmon spawners to Salmon Fishing Areas $\mathbf{1 5}$ to $\mathbf{2 3}$ for Canada and for USA used in the run reconstruction.

| Year of returnto rivers | Spawners of small salmon |  |  |  | SFA 17 | SFA 18 |  |  | SFA 19-21 |  | SFA 23 |  | USA |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SFA 15 |  | SFA 16 |  |  |  |  |  |  |  |  |  |
|  | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max |  | Min | Max | Point estimate |
| 1970 | 1417 | 4396 | 958 | 45876 | 0 | 0 | 167 | 842 | 9429 | 17358 | 3886 | 610 |  |
| 1971 | 1056 | 3277 | 22463 | 38195 | 0 | 0 | 41 | 208 | 7246 | 13339 | 1216 | 2509 | 29 |
| 1972 | 1034 | 3208 | 27639 | 48023 | 0 | 0 | 82 | 416 | 7616 | 14021 | 0 | 1 | 17 |
| 1973 | 1505 | 4668 | 31703 | 51349 | 3 | 7 | 325 | 1645 | 9502 | 17492 | 4037 | 5575 | 13 |
| 1974 | 1098 | 3405 | 57376 | 8975 | 5 | 0 | 118 | 595 | 16680 | 30706 | 8071 | 10777 | 40 |
| 1975 | 1195 | 3707 | 50438 | 7888 | 0 | 0 | 71 | 357 | 5819 | 10712 | 15363 | 20442 | 67 |
| 1976 | 2480 | 7692 | 64526 | 104130 | - 8 | 22 | 188 | 951 | 14196 | 26134 | 17572 | 23601 | 151 |
| 1977 | 2467 | 7653 | 13270 | 25338 | 0 | 0 | 135 | 684 | 15120 | 27835 | 9196 | 12129 | 54 |
| 1978 | 1398 | 4337 | 14689 | 24833 | 3 | 0 | 49 | 248 | 2857 | 5259 | 4256 | 5680 | 127 |
| 1979 | 2104 | 6528 | 31829 | 51876 | - 1 | 4 | 1170 | 5915 | 15716 | 28932 | 11640 | 16801 | 247 |
| 1980 | 2996 | 9293 | 27791 | 44943 | 7 | 18 | 327 | 1655 | 18876 | 34749 | 19597 | 25941 | 722 |
| 1981 | 3183 | 9874 | 35423 | 80370 | 151 | 390 | 1762 | 8908 | 21096 | 38837 | 7805 | 12782 | 1009 |
| 1982 | 3038 | 9027 | 51324 | 106423 | 102 | 263 | 1354 | 6847 | 11244 | 20700 | 6532 | 10357 | 290 |
| 1983 | 820 | 2486 | 13298 | 30045 | 10 | 25 | 133 | 674 | 5653 | 10408 | 5132 | 8454 | 255 |
| 1984 | 1620 | 4971 | 7389 | 28271 | 10 | 25 | 177 | 1200 | 13658 | 25143 | 10290 | 16412 | 540 |
| 1985 | 3557 | 10936 | 32275 | 71106 | 66 | 170 | 145 | 1788 | 18024 | 33181 | 8164 | 13036 | 363 |
| 1986 | 5589 | 16990 | 71918 | 146983 | 330 | 852 | 63 | 1729 | 18187 | 33481 | 10725 | 16634 | 66 |
| 1987 | 4867 | 14920 | 49971 | 104131 | 665 | 1718 | 527 | 3075 | 20213 | 37210 | 10257 | 14561 | 1087 |
| 1988 | 6664 | 20468 | 71967 | 149800 | 899 | 2320 | 344 | 2388 | 18125 | 33366 | 13061 | 19764 | 923 |
| 1989 | 3191 | 9741 | 37696 | 85724 | 233 | 603 | 232 | 1650 | 18973 | 34928 | 13124 | 20066 | 1080 |
| 1990 | 3996 | 12190 | 46902 | 99996 | 1074 | 2771 | 229 | 1750 | 22080 | 40648 | 10025 | 15381 | 617 |
| 1991 | 2215 | 6872 | 39648 | 78522 | 919 | 2371 | 271 | 2068 | 7363 | 13556 | 9495 | 14139 | 235 |
| 1992 | 4426 | 13728 | 116657 | 178949 | 1092 | 2818 | 189 | 1634 | 10125 | 18640 | 9485 | 14326 | 1124 |
| 1993 | 2891 | 8968 | 52050 | 157056 | 745 | 1922 | 261 | 1805 | 9970 | 18354 | 5762 | 6868 | 44 |
| 1994 | 4554 | 14125 | 25649 | 43764 | 118 | 292 | 179 | 1266 | 2661 | 4900 | 4965 | 5738 | 427 |
| 1995 | 1451 | 4501 | 34650 | 53746 | 250 | 375 | 148 | 1055 | 6512 | 11988 | 8025 | 9218 | 213 |
| 1996 | 3017 | 9359 | 19511 | 29260 | - 258 | -387 | 1005 | 5596 | 10909 | 20082 | 11576 | 13892 | 651 |
| 1997 | 2899 | 8991 | 8702 | 15524 | 256 | 384 | 203 | 1290 | 2917 | 5370 | 3971 | 4433 | 365 |
| 1998 | 3144 | 9752 | 14650 | 22573 | 255 | 382 | 228 | 1464 | 8818 | 11912 | 8775 | 10348 | 403 |
| 1999 | 2465 | 7646 | 12265 | 17553 | 253 | 380 | 347 | 1837 | 3895 | 5261 | 5196 | 6048 | 419 |
| 2000 | 3727 | 11560 | 19220 | 27159 | 252 | 378 | 314 | 1717 | 6148 | 8305 | 4455 | 5087 | 270 |
| 2001 | 2470 | 7663 | 16165 | 22854 | 250 | 376 | 403 | 2217 | 2315 | 3127 | 2210 | 2530 | 266 |
| 2002 | 5857 | 18166 | 26637 | 36448 | 249 | 373 | 426 | 2334 | 5180 | 6998 | 3232 | 3689 | 450 |
| 2003 | 1557 | 4829 | 16300 | 24877 | 248 | 371 | 396 | 2201 | 2829 | 3822 | 2069 | 2469 | 237 |
| 2004 | 6043 | 18744 | 28270 | 40876 | 246 | 369 | 496 | 2934 | 3833 | 5178 | 3229 | 4039 | 319 |
| 2005 | 2056 | 6377 | 16882 | 28429 | 246 | 368 | 300 | 1881 | 2854 | 3855 | 3433 | 4450 | 319 |
| 2006 | 4359 | 13522 | 18147 | 31462 | 247 | 370 | 358 | 2201 | 5119 | 6915 | 3528 | 4501 | 450 |
| 2007 | 2127 | 6597 | 15184 | 27735 | 248 | 372 | 330 | 1905 | 4176 | 5642 | 2305 | 2937 | 297 |
| 2008 | 6798 | 21086 | 17040 | 29941 | 249 | 373 | 714 | 5018 | 7252 | 9801 | 5729 | 7467 | 814 |
| 2009 | 2581 | 8007 | 4929 | 10621 | 233 | 350 | 102 | 931 | 2051 | 2773 | 1472 | 1864 | 241 |
| 2010 | 4090 | 12688 | 31127 | 47580 | 256 | 384 | 374 | 2521 | 3674 | 4963 | 9032 | 11901 | 525 |
| 2011 | 5114 | 15864 | 25352 | 48064 | 290 | 435 | 562 | 3558 | 3601 | 4864 | 4391 | 5867 | 1080 |
| 2012 | 2172 |  |  |  |  |  |  |  | 343 | 463 | 167 | 208 |  |
|  | SFisss | SF15SS | SF16SS | SF16ss | SF17ss | SF17ss | SF18ss | SF188s | SF19_21 | SF19_21 | SF23ss | SF23ss | USASSm |
| Winbugs labels | $\mathrm{m}=$ L] | $\mathrm{m}, \mathrm{H}[]$ | $\mathrm{m} \_ \text {ㄴI }$ | m_H] | m_L] | $\mathrm{m}=\mathrm{H}]$ | m_L] | $\mathrm{m}=\mathrm{H}]$ | ssm_L] | SSm_H[] | m_L] | $\mathrm{m}=\mathrm{H}]$ |  |

6.x. Estimated SMALL salmon returns for the six North American regions and North American total from the run reconstruction model.

| Return | Labrador |  | Newfoundland |  |  |  | Quebec |  |  | Gulf |  | Scotia-Fundy |  |  |  | USA |  |  | North America |  | 95th perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Median | 5th | 95th | Median | 5th pe | 95t | Median | 5th | 95t | Medi | 5th | 95th | Median | 5th | 95t | edian | 5 th perc. | 95th perc | Median | 5th pe |  |
| 1970 | 49310 | 34190 | 72900 | 135600 | 120200 | 150900 | 23630 | 19380 | 27880 | 62990 | 53910 | 72010 | 26530 | 22790 | 30310 |  |  |  | 299100 | 272900 | 328600 |
| 1971 | 64350 | 44690 | 95180 | 118800 | 105600 | 132000 | 18720 | 15340 | 22070 | 49860 | 42660 | 56960 | 18850 | 16050 | 21660 | 32 | 32 | 32 | 271400 | 244400 | 305400 |
| 1972 | 48540 | 33710 | 71630 | 110600 | 97640 | 123400 | 15600 | 12780 | 18390 | 62840 | 53690 | 72010 | 16960 | 14070 | 19840 | 18 | 18 | 18 | 255500 | 231500 | 283300 |
| 1973 | 13960 | 9439 | 19810 | 159900 | 142000 | 177700 | 20750 | 17000 | 24660 | 160 | 54160 | 72200 | 24400 | 20750 | 28070 | 23 | 23 | 23 | 282400 | 260800 | 304100 |
| 197 | 54100 | 37550 | 79620 | 120500 | 106700 | 200 | 21000 | 17210 | 24770 | 300 | - 83720 | 12900 | 43610 | 37170 | 980 | 55 | 55 | 56 | 338700 | 09000 | 371400 |
| 1975 | 10300 | 71440 | 1534 | 151000 | 133100 | 8900 | 22570 | 8510 | 650 | 8390 | 75570 | 1200 | 3860 | 30440 | 37270 | -84 | 83 | 85 | 400100 | 8400 | 454700 |
| 1976 | 73760 | 1260 | 09000 | 158600 | 139100 | 178100 | 24950 | 20460 | 330 | 128800 | 110800 | 146800 | 5291 | 4662 | 59180 | 186 | - 184 | 188 | 44080 | 40190 | 484600 |
| 1977 | 65530 | 45660 | - 96880 | 159600 | 00 | 179100 | 770 | 850 | 830 | 60 | 39930 | 52630 | 6140 | 40250 | 52060 | -75 | - 74 | 76 | 341700 | 309700 | 378900 |
| 1978 | 32800 | 22900 | 47940 | 139400 | 121800 | 157000 | 21200 | 17390 | 25030 | 41100 | 36190 | 46020 | 15800 | 14480 | 17130 | 155 | 154 | 156 | 251300 | 228800 | 274900 |
| 1979 | 42370 | 29250 | 62890 | 151800 | 133000 | 170700 | 27100 | 22220 | 31980 | 72320 | 62520 | 82150 | 48850 | 42270 | 55400 | 250 | 248 | 252 | 344000 | 315700 | 373800 |
| 1980 | 96090 | 66250 | 142900 | 172400 | 152200 | 192300 | 37240 | 30530 | 43930 | 63240 | 54520 | 71980 | 70620 | 62700 | 78560 | 818 | 811 | 825 | 441600 | 400200 | 493400 |
| 1981 | 105600 | 72590 | 157700 | 225300 | 197700 | 253400 | 52090 | 42700 | 61470 | 106400 | 85480 | 127400 | 59360 | 51000 | 67740 | 1130 | 1120 | 1140 | 552300 | 497900 | 615000 |
| 1982 | 73340 | 50570 | 109000 | 200700 | 177500 | 224100 | 29630 | 24880 | 34970 | 121300 | 96160 | 146400 | 36050 | 31330 | 40830 | - 334 | 331 | 337 | 463300 | 417800 | 512900 |
| 1983 | 45970 | 31780 | 68210 | 156700 | 137800 | 175600 | 22490 | 440 | 26530 | 37170 | 29620 | 44740 | 22620 | 19830 | 25390 | 295 | 292 | 298 | 286400 | 259100 | 316200 |
| 1984 | 24080 | 16770 | 35620 | 206100 | 179500 | 233000 | 26220 | 23920 | 28520 | 54230 | 44690 | 63830 | 42760 | 36580 | 48920 | - 598 | 593 | 603 | 35500 | 324200 | 386100 |
| 1985 | 43260 | 29840 | 64430 | 195400 | 168300 | 222800 | 28020 | 25540 | 30480 | 86150 | 68180 | 104200 | 47430 | 40140 | 54750 | 392 | 389 | 396 | 402200 | 363700 | 441500 |
| 1986 | 65430 | 100 | 97560 | 200300 | 175000 | 225600 | 3350 | 7340 | 3360 | 161400 | 127200 | 195600 | 49260 | 41640 | 56850 | 758 | 751 | 765 | 519700 | 467700 | 573100 |
| 1987 | 82050 | 56410 | 122500 | 135400 | 118500 | 152400 | 45940 | 42140 | 49730 | 122500 | 97220 | 147400 | 51240 | 43330 | 59220 | 1128 | 1118 | 1138 | 440200 | 395300 | 490800 |
| 1988 | 75720 | 51820 | 113000 | 217300 | 189900 | 247700 | 53080 | 48970 | 57160 | 172600 | 136700 | 208300 | 51900 | 44140 | 59660 | 992 | 983 | 1001 | 573900 | 517400 | 632600 |
| 1989 | 51830 | 3575 | 7721 | 10770 | 94730 | 2000 | 490 | 3800 | 880 | 102800 | 81070 | 124700 | 54620 | 46470 | 62750 | 125 | 124 | 269 | 361200 | 326900 | 397800 |
| 1990 | 30260 | 20880 | 4503 | 152400 | 138200 | 166500 | 47380 | 44000 | 50750 | 117200 | 92840 | 141300 | 55250 | 46450 | 64080 | 687 | 681 | 693 | 40430 | 371100 | 437600 |
| 1991 | 24 | 16590 | 36440 | 105600 | 96390 | 114800 | 37110 | 34530 | 39720 | 84980 | 67320 | 102600 | 28220 | 24520 | 31950 | 310 | 307 | 313 | 281300 | 257900 | 305200 |
| 1992 | 34310 | 24190 | 51160 | 229000 | 199900 | 258000 | 42000 | 38940 | 45060 | 192800 | 164400 | 221200 | 34010 | 29350 | 38640 | 1194 | 1183 | 1205 | 534600 | 490200 | 579200 |
| 1993 | 45800 | 33250 | 66820 | 265500 | 235200 | 295700 | 36390 | 33860 | 38930 | 136000 | 89020 | 183500 | 25720 | 21910 | 29500 | 466 | 462 | 470 | 511600 | 451500 | 572300 |
| 1994 | 33880 | 25150 | 48330 | 161000 | 138800 | 183200 | 34920 | 32520 | 37310 | 67310 | 57200 | 77490 | 10460 | 9358 | 11570 | 436 | 432 | 440 | 309100 | 282000 | 337000 |
| 199 | 47870 | 35880 | - 66930 | 204100 | 173400 | 234600 | 110 | 26200 | 30020 | 60590 | 51900 | 69430 | 20000 | 17470 | 22530 | 13 | 211 | 215 | 362200 | 326500 | 398300 |
| 1996 | 90150 | 780 | 127500 | 313500 | 269400 | 357600 | 37280 | 34800 | 39770 | 55310 | 47250 | 63400 | 31770 | 27490 | 36070 | 651 | 645 | 657 | 531100 | 477900 | 587600 |
| 1997 | 95270 | 73620 | 130900 | 177000 | 159200 | 194900 | 28870 | 26750 | 30990 | 30590 | 24730 | 36390 | 9380 | 8254 | 10500 | 365 | 362 | 368 | 342700 | 311700 | 382000 |
| 1998 | 151200 | 102900 | 199800 | 183800 | 171300 | 196200 | 29380 | 26890 | 31900 | 40840 | 34240 | 47430 | 20380 | 18740 | 22020 | 403 | 399 | 407 | 425900 | 375600 | 47660 |
| 1999 | 147400 | 100100 | 194700 | 201300 | 185700 | 216900 | 31290 | 28780 | 33790 | 35520 | 30790 | 40260 | 10590 | 9819 | 11360 | 419 | 415 | 423 | 426600 | 376000 | 477100 |
| 2000 | 181900 | 123600 | 240100 | 228800 | 216800 | 240700 | 29030 | 25960 | 32140 | 52160 | 45120 | 59220 | 12360 | 11330 | 13380 | 270 | 268 | 272 | 504700 | 444700 | 564300 |
| 2001 | 145400 | 98920 | 192000 | 156300 | 148100 | 164500 | 20150 | 18410 | 21890 | 42430 | 36950 | 47910 | 5417 | 5006 | 5831 | 266 | 264 | 268 | 369900 | 322600 | 417300 |
| 2002 | 102600 | 66300 | 138900 | 155600 | 143300 | 167900 | 32570 | 30300 | 34850 | 70550 | 60790 | 80220 | 9851 | 8996 | 10710 | 450 | 446 | 454 | 371500 | 331400 | 411700 |
| 2003 | 85600 | 51930 | 119000 | 242500 | 232800 | 252100 | 26680 | 24610 | 28730 | 41020 | 35100 | 47040 | 5844 | 5341 | 6348 | 237 | 235 | 239 | 401800 | 366100 | 437500 |
| 2004 | 95000 | 72320 | 117800 | 210100 | 192100 | 228200 | 35990 | 32470 | 39520 | 76030 | 64650 | 87360 | 8394 | 7635 | 9151 | 319 | 316 | 322 | 425900 | 393800 | 458000 |
| 2005 | 220800 | 166300 | 275200 | 221500 | 176600 | 266500 | 24270 | 22080 | 26450 | 45360 | 37450 | 53310 | 7488 | 6788 | 818 | - 319 | - 316 | 322 | 519600 | 446300 | 593000 |
| 2006 | 213700 | 140400 | 286600 | 212800 | 194300 | 231400 | 29750 | 27550 | 31960 | 54380 | 44070 | 64840 | 10270 | 9293 | 11260 | 450 | 446 | 454 | 521300 | 445200 | 597300 |
| 2007 | 194700 | 138200 | 251100 | 183500 | 158700 | 208600 | 22600 | 20600 | 24620 | 42000 | 33460 | 50620 | 7733 | 6979 | 8486 | 297 | 294 | 300 | 451000 | 387600 | 513900 |
| 2008 | 203800 | 148900 | 258600 | 247800 | 222200 | 273300 | 37570 | 34560 | 40590 | 62400 | 50210 | 74610 | 15370 | 13880 | 16860 | 814 | 807 | 821 | 567800 | 504700 | 630700 |
| 2009 | 89670 | 43140 | 135800 | 222600 | 194200 | 250900 | 22230 | 20330 | 24110 | 22980 | 18040 | 27980 | 4241 | 3844 | 4639 | 241 | 239 | 243 | 361700 | 305300 | 418100 |
| 2010 | 91910 | 59550 | 124000 | 267700 | 256200 | 279300 | 28080 | 25680 | 30470 | 76560 | 64580 | 88490 | 14880 | 13410 | 16350 | 525 | 520 | 530 | 479600 | 442400 | 516700 |
| 2011 | 271600 | 97590 | 447100 | 243300 | 216200 | 270300 | 38400 | 35540 | 41270 | 75910 | 59700 | 92110 | 9448 | 8511 | 10390 | 1080 | 1070 |  | 640000 | 463500 | 817500 |
| 2012 | 172800 | 75060 | 270300 | 242300 | 215100 | 269300 | 25240 | 23050 | 27440 | 17670 | 13760 | 21600 | 609 | 550 | 667 | 24 | 24 | 24 | 458500 | 357100 | 56020 |

6.xi. Estimated SMALL salmon spawners for the six North American regions and North American total from the run reconstruction model.

| Spawner L | Labrador |  | Newfoundland |  |  |  | Quebec <br> Median | 5th perc. | 95th perc |  | 5th perc. | Scotia-Fundy |  |  |  | USA | 5th perc. | North America |  |  | 95th perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Median | 5th perc. | 95th perc | Median | 5th perc. | 95th perc |  |  |  |  |  | 95th perc | Median | 5th perc. | 95th perc | Medi |  | 95th perc | Median | 5th perc. |  |
| 1970 | 45290 | 30180 | 68890 | 105200 | 89920 | 120600 | 13820 | 11330 | 16290 | 39330 | 30330 | 48340 | 18400 | 14630 | 22150 |  |  |  |  |  |  |
| 1971 | 60420 | 40750 | 91250 | 92190 | 78940 | 105400 | 11670 | 9571 | 13770 | 32620 | 25500 | 39700 | 12170 | 9347 | 14970 | 29 | 29 | 29 | 209900 | 183000 | 244000 |
| 1972 | 4559 | 30 | 68880 | 86200 | 733 | 9906 | 10270 | 8423 | 12120 | 40250 | 1020 | 49350 | 10820 | 7939 | 300 | 17 | 17 | 17 | 4100 | 200 | 221700 |
| 1973 | 6463 | 1947 | 12320 | 124300 | 106500 | 142200 | 13730 | 11270 | 16210 | 45590 | 36630 | 54570 | 18310 | 14650 | 21950 | 13 | ${ }^{13}$ | 13 | 208700 | 187400 | 230100 |
| 1974 | 51600 | 35050 | 77120 | 94080 | 80410 | 107700 | 12590 | 10320 | 14850 | 76140 | 61610 | 90770 | 33100 | 26690 | 39530 | 40 | 40 | 40 | 268800 | 239500 | 301300 |
| 1975 | 99080 | 67470 | 149500 | 117500 | 99710 | 135500 | 14510 | 11900 | 17120 | 67310 | 54530 | 80130 | 26160 | 22770 | 29570 | 67 | 66 | 68 | 325800 | 284200 | 380400 |
| 1976 | 68030 | 45530 | 103300 | 124100 | 104400 | 143700 | 16210 | 13310 | 19140 | 90030 | 72140 | 107900 | 40770 | 34470 | 47020 | 151 | 150 | 152 | 340900 | 302100 | 384800 |
| 1977 | 60930 | 41070 | 92290 | 125200 | 105800 | 144700 | 15010 | 12300 | 17700 | 24770 | 18670 | 30900 | 32160 | 26250 | 38020 | - 54 | - 54 | 54 | 259600 | 227800 | 296600 |
| 1978 | 30100 | 20210 | 45250 | 110800 | 93190 | 128400 | 14300 | 11730 | 16870 | 22770 | 17970 | 27600 | 9019 | 7699 | 10350 | 127 | 126 | 128 | 188100 | 165700 | 211300 |
| 1979 | 38250 | 25130 | 58770 | 120800 | 101800 | 139600 | 19840 | 16270 | 23400 | 49710 | 40130 | 59270 | 36560 | 29950 | 43130 | 247 | 245 | 249 | 266600 | 23860 | 296300 |
| 1980 | 92290 | 62450 | 139100 | 136500 | 116500 | 156600 | 26030 | 21320 | 700 | 3540 | 5040 | 51940 | 49570 | 41610 | 57510 | 722 | 716 | 729 | 349900 | 309100 | 200 |
| 1981 | 100400 | 67400 | 152500 | 178900 | 151000 | 206600 | 10 | 31740 | 5660 | 70080 | 9320 | 9690 | 40280 | 31900 | 48650 | 1009 | 1000 | 1018 | 431600 | 377500 | 9380 |
| 1982 | 69230 | 46470 | 104900 | 158800 | 135500 | 182200 | 21090 | 17290 | 24890 | 89050 | 64200 | 114200 | 24440 | 19710 | 29170 | 290 | 287 | 293 | 364900 | 319400 | 414200 |
| 1983 | 41600 | 27410 | 63840 | 124300 | 105300 | 143100 | 15020 | 12320 | 17740 | 23730 | 16220 | 31280 | 14820 | 12040 | 17600 | 255 | 253 | 257 | 220800 | 193600 | 250500 |
| 1984 | 21140 | 13830 | 32690 | 167100 | 140500 | 193800 | 20370 | 18080 | 22690 | 21820 | 12340 | 31280 | 32760 | 26590 | 38900 | 540 | 535 | 545 | 264600 | 233800 | 295500 |
| 1985 | 40160 | 26740 | 1330 | 158900 | 131700 | 186200 | 20090 | 17630 | 22560 | 59940 | 42290 | 77810 | 36190 | 28890 | 43460 | 363 | 360 | 366 | 317200 | 278700 | 356300 |
| 1986 | 61960 | 4164 | 94090 | 1627 | 137100 | 188100 | 27710 | 24750 | 30680 | 122300 | 88290 | 156100 | 39500 | 31910 | 47130 | 660 | -654 | 666 | 417100 | 365100 | 470600 |
| 1987 | 766 | 5104 | 117100 | 11100 | 940 | 127900 | 32780 | 29040 | 36530 | 89960 | 65170 | 114800 | 41120 | 33170 | 49110 | 108 | 107 | 97 | 354600 | 310000 | 404900 |
| 1988 | 70200 | 46300 | 107500 | 177500 | 150400 | 204700 | 36360 | 32320 | 40430 | 127200 | 91860 | 163000 | 42190 | 34450 | 49970 | 923 | 915 | 931 | 456800 | 400900 | 515500 |
| 1989 | 47150 | 31070 | 72530 | 89160 | 76270 | 101900 | 30720 | 27530 | 33870 | 69500 | 47840 | 91230 | 43570 | 35460 | 51620 | 1080 | 1070 | 1090 | 282800 | 248300 | 319200 |
| 1990 | 26950 | 17570 | 41720 | 122400 | 108200 | 136600 | 32790 | 29470 | 36130 | 84490 | 60320 | 108600 | 44050 | 35250 | 52850 | 617 | 611 | 623 | 312200 | 279400 | 345700 |
| 1991 | 21930 | 14270 | 34120 | 85070 | 75840 | 94300 | 25240 | 22670 | 27770 | 66430 | 48860 | 84000 | 22300 | 18560 | 26000 | 235 | 233 | 237 | 222000 | 198800 | 245900 |
| 1992 | 31550 | 21430 | 48390 | 205200 | 176200 | 234400 | 27380 | 24350 | 30380 | 159800 | 131600 | 187900 | 26270 | 21640 | 30940 | 1124 | 111 | 1134 | 452700 | 408500 | 497300 |
| 1993 | 43110 | 30560 | 64140 | 2392 | 2089 | 269400 | 22010 | 520 | 490 | 112700 | 65610 | 160100 | 20 | 16 | 24260 | 444 | 440 | 448 | 43970 | 379400 | 500300 |
| 1994 | 30960 | 22230 | 45420 | 129900 | 107500 | 152100 | 20720 | 18380 | 23070 | 45030 | 35330 | 54710 | 9133 | 8046 | 10220 | 427 | 423 | 431 | 237200 | 210400 | 264900 |
| 1995 | 45040 | 33060 | 64110 | 171200 | 140500 | 201900 | 17700 | 15830 | 19590 | 48110 | 39410 | 56770 | 17870 | 15340 | 20390 | 213 | 211 | 215 | 301500 | 266000 | 337800 |
| 1996 | 87210 | 64840 | 124500 | 274700 | 230900 | 318800 | 23180 | 20750 | 25600 | 34190 | 28250 | 40130 | 28220 | 23960 | 32530 | 651 | 645 | 657 | 450700 | 397900 | 507000 |
| 1997 | 92680 | 71040 | 128300 | 151900 | 134000 | 169700 | 17960 | 15890 | 20030 | 19110 | 14680 | 23560 | 8344 | 7222 | 9468 | 365 | 362 | 368 | 291400 | 260700 | 330600 |
| 199 | 14860 | 100400 | 300 | 5400 | 145900 | 7800 | 190 | 00 | 380 | 26250 | 1240 | 31200 | 19920 | 18280 | 21550 | 403 | 399 | 40 | 37470 | 324400 | 425500 |
| 1999 | 144900 | 97570 | 192200 | 176400 | 160700 | 192100 | 23730 | 21240 | 26220 | 21380 | 17730 | 25010 | 10200 | 9434 | 10970 | 419 | 415 | 423 | 377000 | 326500 | 427400 |
| 2000 | 178600 | 120400 | 236800 | 204700 | 192700 | 216700 | 21070 | 18000 | 24130 | 32160 | 26740 | 37620 | 12000 | 10970 | 13030 | 270 | 268 | 272 | 44880 | 389200 | 508500 |
| 2001 | 142800 | 96400 | 189500 | 133500 | 125300 | 141700 | 13670 | 12130 | 15220 | 26200 | 22040 | 30350 | 5089 | 4687 | 5497 | 266 | 264 | 268 | 321600 | 274500 | 369100 |
| 2002 | 99990 | 63720 | 136300 | 132900 | 120600 | 145200 | 21350 | 19130 | 23560 | 45260 | 37640 | 52930 | 9549 | 8699 | 10400 | 450 | 446 | 454 | 309400 | 27000 | 349100 |
| 2003 | 83000 | 49330 | 116400 | 219600 | 209900 | 229200 | 19310 | 17270 | 21360 | 25420 | 21070 | 29730 | 5590 | 5096 | 6091 | 237 | 235 | 239 | 353200 | 317600 | 388400 |
| 2004 | 92590 | 69910 | 115400 | 188400 | 170500 | 206500 | 26300 | 22800 | 29790 | 49050 | 40310 | 57720 | 8140 | 7389 | 8888 | 319 | 316 | 322 | 364900 | 333500 | 396200 |
| 2005 | 218000 | 163500 | 272500 | 197200 | 152100 | 241900 | 18280 | 16140 | 20440 | 28240 | 22530 | 34000 | 7296 | 6608 | 7989 | 319 | 316 | 322 | 469000 | 396100 | 542300 |
| 2006 | 211500 | 138200 | 284400 | 191000 | 172400 | 209600 | 21600 | 19420 | 23780 | 35310 | 27540 | 43110 | 10030 | 9066 | 11000 | 450 | 446 | 454 | 469800 | 393800 | 545400 |
| 2007 | 192500 | 136000 | 248900 | 167700 | 142800 | 192800 | 16720 | 14720 | 18700 | 27270 | 21040 | 33480 | 7527 | 6786 | 8276 | 297 | 294 | 300 | 412000 | 348800 | 474900 |
| 2008 | 201200 | 146400 | 256100 | 217400 | 191900 | 243000 | 26700 | 23730 | 29680 | 40620 | 31100 | 50100 | 15120 | 13650 | 16600 | - 814 | 207 | 821 | 501900 | 439400 | 564500 |
| 2009 | 87980 | 41450 | 134100 | 197100 | 168700 | 225600 | 16220 | 14350 | 18090 | 13880 | 10060 | 17690 | 4080 | 3692 | 4469 | 241 | 239 | 243 | 319400 | 263100 | 375600 |
| 2010 | 89960 | 57600 | 122100 | 235300 | 223700 | 246900 | 20500 | 18140 | 22870 | 49530 | 40650 | 58360 | 14780 | 13310 | 16250 | 525 | 520 | 530 | 410500 | 374500 | 446400 |
| 2011 | 269500 | 95430 | 445000 | 214200 | 187200 | 241100 | 27810 | 24970 | 30650 | 49630 | 37820 | 61480 | 9362 | 8419 | 10300 | 1080 | 1070 | 1090 | 571100 | 396000 | 748500 |
| 2012 | 170700 | 72940 | 268200 | 213800 | 186700 | 241200 | 18410 | 16240 | 20560 | 10650 | 7616 | 13680 | 590 | 532 | 649 | 24 | 24 | 24 | 414200 | 312600 | 515600 |

6.xii. Estimated LARGE salmon returns for the six North American regions and North American total from the run reconstruction model.

| Return | Labrador |  | Newfoundland |  |  |  | Quebec |  |  | Gulf |  | Scotia-Fundy |  |  |  | USA |  | North America |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Median | 5th perc. | 95th perc N | Median | 5th perc. | 95th p | Median | 5th perc. | 95th perc N | Median | 5th perc. | 95th perc N | Median | 5th per | 95th perc N | Media | erc. | 95th perc | Median | 5th perc. | 95th per |
| 1970 | 10120 | 4969 | 17000 | 14870 | 11820 | 17890 | 103400 | 84770 | 122000 | 69570 | 67140 | 71990 | 20280 | 17980 | 22610 |  |  |  | 218600 | 198200 | 238900 |
| 1971 | 14410 | 7068 | 24290 | 12570 | 10010 | 15120 | 59160 | 48520 | 820 | 40050 | 37600 | 42490 | 1588 | 14120 | 50 | 653 | 647.1 | 658.9 | 143200 | 128300 | 80 |
| 1972 | 12440 | 6098 | 20850 | 12670 | 10110 | - 15220 | 77220 | 63310 | 91110 | 56970 | 48950 | 65050 | 18980 | 17110 | 20840 | 1383 | 1371 | 1395 | 180100 | 161300 | 80 |
| 1973 | 17300 | - 8494 | 29170 | 17330 | 13780 | 20900 | 85200 | 9920 | 100600 | 53370 | 45550 | 61170 | 1476 | 1343 | 16090 | 1427 | 1414 | 1440 | 190000 | 168700 | O |
| 1974 | 17060 | - 8365 | 28790 | 260 | 12680 | 15850 | 114200 | 93730 | 4900 | 7610 | 5910 | 260 | 28580 | 26300 | 30830 | 1394 | 1381 | 1407 | 253700 | 226900 | 0 |
| 1975 | 15890 | 7810 | 26830 | 18420 | 16110 | 20710 | 97070 | 79580 | 114600 | 50360 | 43020 | 57760 | 30630 | 28020 | 33230 | 2331 | 2310 | 2352 | 215200 | 192800 | 237800 |
| 1976 | 18320 | 9003 | 30800 | 16640 | 14640 | 850 | 96450 | 79160 | 113900 | 48750 | 41370 | 56130 | 28800 | 25970 | 31630 | 1317 | 1305 | 1329 | 210900 | 188100 | 34100 |
| 1977 | 16250 | 7979 | 27390 | 600 | 2950 | 6250 | 113800 | 93350 | 134300 | 87840 | 75210 | 0400 | 8080 | 34600 | 41540 | 199 | 1980 | 2016 | 273000 | 246000 | 300400 |
| 1978 | 12740 | 6259 | 145 | 1135 | 10340 | 12340 | 102400 | 84000 | 120900 | 43840 | 38800 | 4893 | 22260 | 20560 | 23960 | 420 | 4170 | 246 | 197300 | 176100 | 218500 |
| 1979 | 7299 | - 3569 | 12250 | 198 | 6297 | 100 | 56530 | 46310 | 66690 | 17850 | 15670 | 20030 | 12810 | 11600 | 14030 | 1942 | 1925 | 195 | 103800 | 92130 | 0 |
| 1980 | 年60 | -8521 | 320 | 50 | 11110 | 90 | 34400 | 110200 | 158600 | 60 | 520 | 90 | 43730 | 580 | 789 | 5796 | 5744 | 5848 | 276400 | 247700 | 00 |
| 1981 | 15620 | 7668 | 26300 | 8880 | 3300 | 32450 | 105600 | 86560 | 124600 | 39320 | 32940 | 45710 | 28210 | 25460 | 30980 | 5601 | 5551 | 565 | 223700 | 200400 | 247100 |
| 1982 | 11590 | 5670 | 19530 | 1600 | 10990 | 13110 | 93720 | 76770 | 110500 | 54130 | 42800 | 65380 | 23670 | 21520 | 25810 | 6056 | 6002 | 6111 | 201000 | 17820 | 223900 |
| 1983 | 8353 | 4112 | 4070 | 2460 | 11280 | 13640 | 76860 | 63030 | 90670 | 40630 | 33740 | 47590 | 20600 | 18380 | 22820 | 2155 | 2136 | 2174 | 161400 | 144200 | 178500 |
| 1984 | 6028 | 2957 | 10110 | 12390 | 50 | 15630 | 71110 | 67670 | 74520 | 32750 | 23460 | 41990 | 24530 | 21150 | 27860 | 3222 | 3193 | 3251 | 150200 | 138200 | 162100 |
| 1985 | 4715 | 2318 | 7958 | 2990 | 7681 | 180 | 73540 | 69250 | 77870 | 580 | 1930 | 57200 | 34180 | 29340 | 39030 | 5529 | 5479 | 557 | 173700 | 158500 | 188700 |
| 1986 | 8162 | 4010 | 90 | 2300 | 9452 | 50 | 90 | 880 | 140 | 760 | 300 | 88040 | 28250 | 23810 | 32680 | 6176 | 6120 | 6231 | 211300 | 189800 | 232800 |
| 1987 | 11030 | - 5423 | 18570 | 8435 | 6454 | 420 | 82930 | 78650 | 190 | 46660 | 34170 | 59070 | 1770 | 1502 | 0360 | 3081 | 3053 | 3109 | 170200 | 1548 | 5700 |
| 1988 | 6916 | - 3377 | 1520 | 12980 | 9886 | 16080 | 90600 | 85440 | 95730 | 33700 | 39590 | 67680 | 16440 | 13710 | 19170 | 3286 | 325 | 3316 | 184100 | 167800 | 200400 |
| 1989 | 6651 | 3256 | 11170 | 6915 | 587 | 8446 | 81310 | 77140 | 85460 | 42570 | 31500 | 53710 | 18500 | 15630 | 21390 | 319 | 3168 | 3226 | 159300 | 146200 | 172600 |
| 1990 | 3824 | 1875 | 6443 | 10270 | 8359 | 12200 | 79860 | 75120 | 84600 | 56710 | 39810 | 73450 | 16020 | 13500 | 18540 | 5051 | 5005 | 5096 | 171900 | 153600 | 189900 |
| 1991 | 1877 | 920.5 | 3148 | 7567 | 6154 | 8987 | 73690 | 69470 | 77910 | 57550 | 4022 | 75240 | 15650 | 13430 | 17870 | 2647 | 2623 | 2671 | 159100 | 140600 | 17760 |
| 1992 | 7534 | 3982 | 12740 | 31550 | 22170 | 850 | 120 | 69690 | 78560 | 60110 | 51470 | 68790 | 14280 | 12310 | 16260 | 2459 | 2437 | 248 | 190400 | 175900 | 204800 |
| 1993 | 9464 | 5911 | 15070 | 17120 | 13800 | 20430 | 57200 | 54830 | 59570 | 63890 | 34830 | 93130 | 10060 | 8898 | 1210 | 2231 | 2211 | 2251 | 160500 | 3400 | 300 |
| 1994 | 12950 | 81 | 330 | 7360 | 3820 | 2093 | 58130 | 55780 | 60480 | 41430 | 33220 | 49560 | 6312 | 5657 | 6969 | 134 | 133 | 135 | 138000 | 127100 | 149400 |
| 1995 | 25550 | 18140 | 37410 | 19050 | 14690 | 23410 | 67090 | 64440 | 69720 | 48320 | 41270 | 55360 | 7507 | 6588 | 8423 | 1748 | 1732 | 1764 | 169800 | 157400 | 184200 |
| 1996 | 18830 | 13380 | 27640 | 28920 | 23750 | 34090 | 61130 | 58330 | 63940 | 41330 | 33160 | 49550 | 10880 | 9586 | 12170 | 2407 | 2385 | 2429 | 164000 | 151900 | 176900 |
| 1997 | 16190 | 11580 | 23740 | 27980 | 22950 | 33550 | 50320 | 48040 | 52610 | 36020 | 28550 | 43550 | 5585 | 4996 | 6169 | 1611 | 1597 | 1626 | 138200 | 127300 | 149700 |
| 1998 | 13460 | 7982 | 18890 | 35260 | 27430 | 43080 | 38490 | 36330 | 40660 | 30850 | 24450 | 37280 | 3848 | 3535 | 4160 | 1526 | 1512 | 154 | 123400 | 111500 | 135300 |
| 1999 | 16080 | 9547 | 22610 | 32080 | 24940 | 39280 | 40500 | 38160 | 42830 | 27950 | 23030 | 32790 | 4942 | 4593 | 5289 | 1168 | 115 | 1178 | 122700 | 111300 | 134100 |
| 2000 | 21990 | 13030 | 30870 | 27000 | 22990 | 31040 | 38890 | 35900 | 41900 | 30560 | 25220 | 35930 | 2871 | 2614 | 3130 | 533 | 528.2 | 537.8 | 121800 | 110000 | 133700 |
| 2001 | 23210 | 13800 | 32620 | 17870 | 15160 | 20550 | 40700 | 37620 | 43790 | 40650 | 35190 | 46110 | 4657 | 4265 | 5056 | 797 | 789.8 | 804.2 | 127900 | 116000 | 139800 |
| 2002 | 16940 | 9866 | 23980 | 16800 | 13690 | 19910 | 29200 | 26770 | 31630 | 24240 | 19840 | 28600 | 1585 | 1444 | 1725 | 526 | 521.3 | 530.7 | 89270 | 79800 | 98710 |
| 2003 | 14230 | 422 | 20930 | 2460 | 19410 | 29540 | 45420 | 42070 | 48810 | 40740 | 33710 | 47700 | 3526 | 3188 | 3862 | 1199 | 1188 | 1210 | 129600 | 117900 | 1412 |
| 2004 | 17090 | 11590 | 22500 | 22200 | 16980 | 27360 | 39660 | 37000 | 42300 | 41000 | 32740 | 49250 | 3095 | 2825 | 3368 | 1316 | 1304 | 1328 | 124300 | 112700 | 135900 |
| 2005 | 20940 | 12150 | 29830 | 28430 | 20520 | 36300 | 38300 | 35910 | 40670 | 38740 | 30830 | 46660 | 2024 | 1835 | 2212 | 994 | 985 | 1003 | 129400 | 114700 | 144200 |
| 2006 | 21070 | 13280 | 28890 | 35680 | 29990 | 41470 | 35850 | 33540 | 38160 | 38270 | 3070 | 45800 | 2987 | 2683 | 3291 | 1030 | 1021 | 1039 | 134900 | 122100 | 147700 |
| 2007 | 21870 | 12890 | 30870 | 29630 | 23430 | 35800 | 32760 | 30530 | 34990 | 35830 | 29630 | 42090 | 1595 | 1453 | 1737 | 958 | 949.4 | 966.6 | 122600 | 109600 | 135700 |
| 2008 | 26190 | 15850 | 36500 | 28890 | 22550 | 35190 | 38680 | 35430 | 41910 | 29070 | 22770 | 35380 | 3270 | 2920 | 3621 | 1799 | 1783 | 1815 | 127900 | 113500 | 142300 |
| 2009 | 39330 | 20720 | 57980 | 34410 | 23900 | 44970 | 37630 | 35120 | 40130 | 36340 | 30060 | 42620 | 3144 | 2849 | 3440 | 2095 | 2076 | 2114 | 152900 | 129900 | 176100 |
| 2010 | 13820 | 8126 | 19480 | 35350 | 28690 | 41980 | 40070 | 37430 | 42680 | 35280 | 28590 | 41940 | 2513 | 2283 | 2744 | 1098 | 1088 | 1108 | 128100 | 116600 | 139600 |
| 2011 | 43680 | 12370 | 75170 | 43420 | 31290 | 55510 | 51240 | 48090 | 54350 | 71770 | 53960 | 8970 | 4794 | 4317 | 5272 | 3087 | 3059 | 3115 | 218000 | 178600 | 257700 |
| 2012 | 34010 | 12510 | 55510 | 39940 | 29070 | 50780 | 37050 | 34590 | 39510 | 27220 | 21510 | 32990 | 1307 | 1171 | 1445 | 915 | 906.8 | 923.2 | 140400 | 115100 | 16590 |

6.xiii. Estimated LARGE salmon spawners for the six North American regions and North American total from the run reconstruction model.

| Spawner | Labrador |  |  | Newfoundland |  |  | Quebec |  |  | Gulf |  |  | Scotia-Fundy |  |  | USA |  | 95th perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Median | 5th perc. | 95th perc | Median | 5th perc. | 95th perc | Median | 5th perc. | 95th perc | Median | 5th perc. | 95th perc N | Median | 5th perc. | 95th per | Median | 5th perc. |  |
| 1970 | 9563 | 4407 | 16440 | 12740 | 9711 | 15790 | 39140 | 32090 | 46170 | 11900 | 9635 | 14150 | 7888 | 5567 | 10200 |  |  |  |
| 1971 | 13930 | 6582 | 23800 | 10980 | 8435 | 13520 | 20230 | 16590 | 23890 | 11820 | 9427 | 14220 | 8191 | 6429 | 9960 | 490 | 486 | 494 |
| 1972 | 12010 | 5674 | 20430 | 11280 | 8722 | 13830 | 39680 | 32520 | 46810 | 33310 | 25470 | 41150 | 11980 | 10120 | 13850 | 103 | 9 | 1047 |
| 1973 | 16290 | 7485 | 28160 | 15420 | 11860 | 18980 | 40370 | 33080 | 47620 | 35410 | 27770 | 42990 | 7607 | 627 | 89 | 1100 | 1090 | 1110 |
| 1974 | 16250 | 7562 | 27980 | 13050 | 11480 | 14620 | 49090 | 40250 | 57880 | 55820 | 44400 | 67260 | 15200 | 12940 | 1748 | 114 | 113 | 115 |
| 1975 | 15560 | - 7483 | 500 | 170 | 870 | 440 | 800 | 340 | 130 | 3740 | 5450 | 40950 | 17850 | 15250 | 20460 | 1942 | 5 | 960 |
| 1976 | 17490 | - 8173 | 29970 | 15590 | 580 | 17610 | 370 | 810 | 45750 | 29200 | 22140 | 36240 | 6980 | 4150 | 1980 | 112 | 1116 | 136 |
| 1977 | 14970 | 6693 | 26110 | 11850 | 10190 | 13510 | 55870 | 45780 | 65850 | 55590 | 43250 | 67910 | 21550 | 18110 | 25010 | 643 | 637 | 649 |
| 1978 | 11970 | 5492 | 20680 | 9786 | 8780 | 10780 | 51140 | 42000 | 60390 | 19390 | 14620 | 24180 | 10870 | 9160 | 12570 | 3314 | 3284 | 3344 |
| 1979 | 6690 | 2960 | 11650 | 6636 | 5733 | 7542 | 21940 | 17990 | 25880 | 8798 | 6674 | 10890 | 7939 | 6711 | 915 | 150 | 149 | 152 |
| 1980 | 16470 | 7632 | 28430 | 10130 | 9188 | 11070 | 61010 | 49970 | 71900 | 34460 | 26870 | 41980 | 23920 | 19760 | 2806 | 426 | 422 | 301 |
| 1981 | 15100 | 48 | 25780 | 27500 | 23940 | 31060 | 750 | 36680 | 52810 | 16060 | 9831 | 22250 | 12730 | 9968 | 15480 | 4334 | 4295 | 373 |
| 1982 | 10970 | - 5049 | 18900 | 10350 | 8850 | 11860 | 45360 | 190 | 53530 | 27020 | 15760 | 38330 | 10400 | 8262 | 12530 | 464 | 460 | 4685 |
| 1983 | 7925 | 3684 | 13650 | 11080 | 9899 | 12260 | 29680 | 24300 | 34980 | 18060 | 11190 | 24900 | 5721 | 3517 | 7952 | 1769 | 1753 | 785 |
| 1984 | 5518 | 2447 | 9596 | 11870 | 8628 | 15100 | 37110 | 34120 | 40880 | 28510 | 19190 | 37800 | 20020 | 16660 | 23360 | 2547 | 2524 | 2570 |
| 1985 | 4421 | 2024 | 7664 | 10910 | 7658 | 14180 | 35440 | 31560 | 39330 | 43290 | 30640 | 55950 | 28530 | 23670 | 33380 | 4884 | 4840 | 4928 |
| 1986 | 7695 | 3543 | 13230 | 12220 | 9387 | 15080 | 40610 | 36600 | 44680 | 66470 | 47180 | 85780 | 24880 | 20460 | 29340 | 5570 | 5520 | 5620 |
| 1987 | 10390 | 4790 | 17940 | 8391 | 6404 | 10370 | 36050 | 32580 | 39500 | 43910 | 31580 | 56330 | 16060 | 13410 | 1873 | 278 | 275 | 2806 |
| 1988 | - 6206 | 2667 | 10910 | 12910 | 9795 | 16010 | 43160 | 38590 | 47740 | 51950 | 37910 | 65970 | 1479 | 1208 | 1751 | 303 | 30 | 3065 |
| 1989 | 6190 | 2795 | 10710 | 6888 | 5359 | 8423 | 41150 | 37470 | 48830 | 40700 | 29600 | 51740 | 18100 | 15210 | 21010 | 2800 | 2775 | 282 |
| 1990 | 3467 | 1518 | 6086 | 10230 | 8310 | 12140 | 40920 | 36550 | 45260 | 54840 | 38050 | 71620 | 15250 | 12740 | 17760 | 4356 | 4317 | 4395 |
| 1991 | 1784 | 828 | 3055 | 7538 | 6118 | 8953 | 33070 | 29390 | 36740 | 56220 | 38640 | 73860 | 14120 | 11900 | 16330 | 2416 | 2394 | 2438 |
| 1992 | 6752 | 3200 | 11950 | 31410 | 22070 | 40730 | 32360 | 28480 | 36230 | 58200 | 49590 | 66850 | 12980 | 11010 | 14950 | 2292 | 2271 | 2313 |
| 1993 | 9077 | 5524 | 14690 | 16940 | 13610 | 20270 | 24960 | 23130 | 26770 | 62900 | 33810 | 92200 | 8760 | 7601 | 992 | 206 | 20 | 84 |
| 1994 | 12460 | 7991 | 19840 | 10 | 13350 | 20470 | 24460 | 22700 | 220 | 340 | 32210 | 4849 | 5431 | 479 | 6078 | 13 | 133 | 源 |
| 1995 | 25090 | 17680 | 36950 | 18580 | 14210 | 22960 | 34610 | 32670 | 36530 | 4760 | 40590 | 54610 | 7102 | 6186 | 8011 | 1748 | 1732 | 1764 |
| 1996 | 18450 | 13000 | 27260 | 28350 | 23190 | 33530 | 30050 | 27830 | 32250 | 40190 | 32060 | 48260 | 9961 | 8667 | 11250 | 2407 | 2385 | 242 |
| 1997 | 15980 | 11370 | 23530 | 27560 | 22520 | 32620 | 24830 | 23040 | 26630 | 34700 | 27340 | 42100 | 4904 | 4317 | 5491 | 1611 | 1597 | 1626 |
| 1998 | 13150 | 7670 | 18570 | 34930 | 27090 | 42780 | 23040 | 21230 | 24820 | 29920 | 23620 | 36270 | 3474 | 3162 | 3786 | 1526 | 1512 | 1540 |
| 1999 | 1566 | - 9134 | 22200 | 31760 | 24610 | 38920 | 27910 | 25750 | 30080 | 26420 | 21620 | 31230 | 4443 | 4098 | 479 | 1168 | 7 | 1178 |
| 2000 | 21580 | 12620 | 30460 | 26490 | 22490 | 30540 | 26720 | 23850 | 29600 | 29480 | 24210 | 34780 | 2649 | 2392 | 2908 | 1587 | 1573 | 1601 |
| 2001 | 22720 | 13310 | 32140 | 17500 | 14790 | 20190 | 27470 | 24870 | 30080 | 39260 | 33850 | 44660 | 4362 | 3968 | 4756 | 1491 | 1478 | 15 |
| 2002 | 16640 | 9562 | 23680 | 16500 | 13390 | 19610 | 20710 | 18440 | 23010 | 23300 | 18980 | 27650 | 1373 | 1236 | 1511 | 511 | 506 | 516 |
| 2003 | 13880 | 7067 | 20570 | 24090 | 19060 | 29160 | 33760 | 30500 | 37050 | 39460 | 32520 | 46410 | 3292 | 2956 | 3629 | 1192 | 1181 | 120 |
| 2004 | 16680 | 11180 | 22090 | 21840 | 16640 | 27020 | 28140 | 25630 | 30670 | 39620 | 31440 | 47720 | 2961 | 2692 | 3230 | 1283 | 1271 | 129 |
| 2005 | 20520 | 11730 | 29410 | 27870 | 19990 | 35790 | 28080 | 25820 | 30360 | 37300 | 29440 | 45060 | 1900 | 1714 | 2084 | 1088 | 1078 | 1098 |
| 2006 | 20730 | 12940 | 28550 | 35250 | 29500 | 40970 | 26070 | 23920 | 28220 | 36860 | 29410 | 44340 | 2813 | 2513 | 3112 | 1419 | 1406 | 143 |
| 2007 | 21520 | 12530 | 30510 | 29270 | 23090 | 35420 | 23580 | 21460 | 25680 | 34460 | 28330 | 40610 | 1468 | 1329 | 1607 | 1189 | 1178 | 120 |
| 2008 | 25850 | 15500 | 36150 | 28290 | 21950 | 34650 | 29820 | 26650 | 32970 | 27740 | 21530 | 33950 | 3161 | 2811 | 3509 | 2231 | 2211 | 225 |
| 2009 | 38990 | 20380 | 57650 | 34130 | 23610 | 44660 | 28710 | 26290 | 31140 | 34920 | 28690 | 41160 | 3005 | 2713 | 3296 | 2318 | 2297 | 2339 |
| 2010 | 13510 | 7821 | 19180 | 34800 | 28160 | 41470 | 32030 | 29480 | 34590 | 33780 | 27180 | 40390 | 2365 | 2135 | 2595 | 1502 | 1488 | 1515 |
| 2011 | 43460 | 12160 | 74950 | 42840 | 30730 | 54930 | 40260 | 37220 | 43280 | 69980 | 52220 | 87550 | 4709 | 4228 | 5181 | 3914 | 3879 | 3949 |
| 2012 | 33870 | 12360 | 55360 | 39540 | 28650 | 50400 | 28410 | 26050 | 30770 | 26220 | 20530 | 31900 | 1247 | 1110 | 1385 | 2056 | 2038 | 2074 |

6.xiv. Estimated 2SW salmon returns for the six North American regions and North American total from the run reconstruction model.

| Return | Labrador |  |  | Newfound | dland |  | Quebec |  |  | Gulf |  |  | Scotia-Fund |  |  | USA |  |  | North Ame | erica |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Median | 5th perc. | 95th perc | Median | 5th perc. | 95th perc | Median | 5th perc. | 95th perc | Median | 5th perc. | 95th perc N | Median | 5th perc. | 95th perc | Media | 5th perc. | 95th perc N | Median | 5th perc. | 95th perc |
| 1970 | 10120 | 4969 | 17000 | 4133 | 3084 | 5181 | 75470 | 61880 | 89020 | 59590 | 57550 | 61630 | 17130 | 15030 | 19250 |  |  |  | 166800 | 151200 | 182400 |
| 1971 | 14410 | 7068 | 24290 | 3582 | 2606 | 4567 | 43190 | 35420 | 50970 | 34810 | 32630 | 37010 | 13520 | 11880 | 15140 | 653 | 647 | 659 | 110500 | 98260 | 迷 |
| 1972 | 1240 | - 6098 | 2085 | 3735 | 2718 | 4746 | 56370 | 46220 | 66510 | 49460 | 42400 | 56440 | 1599 | 1428 | 17690 | , | 1371 | 1395 | 13970 | 124600 | 155100 |
| 1973 | 17300 | 8494 | 29170 | 4612 | 3467 | 5771 | 62200 | 51040 | 73420 | 47670 | 40640 | 54690 | 12910 | 11690 | 14140 | 1427 | 1414 | 1440 | 146600 | 129200 | 16470 |
| 1974 | 17060 | 8365 | 28790 | 3636 | 2860 | 4417 | 83380 | 68420 | 98460 | 67150 | 57030 | 77350 | 27120 | 24880 | 29370 | 1394 | 1381 | 1407 | 200300 | 178500 | 2500 |
| 1975 | 15890 | 7810 | 26830 | 5210 | 3871 | 6533 | 70860 | 58090 | 83630 | 42950 | 36640 | 49310 | 28870 | 26280 | 31450 | 2331 | 2310 | 2352 | 166600 | 148600 | 185200 |
| 1976 | 18320 | 9003 | 30800 | 4363 | 3312 | 5398 | 70410 | 57780 | 83110 | 40240 | 34240 | 46280 | 26630 | 23830 | 29440 | 1317 | 1305 | 1329 | 161800 | 143200 | 181100 |
| 1977 | 16250 | 7979 | 27390 | 3549 | 2861 | 4238 | 83050 | 68140 | 98010 | 80560 | 68950 | 92180 | 32290 | 28930 | 35660 | 1998 | 1980 | 216 | 218200 | 195700 | 241000 |
| 1978 | 12740 | 6259 | 21450 | 3588 | 2929 | 4252 | 74760 | 61320 | 88250 | 36300 | 32180 | 40430 | 18780 | 17170 | 20390 | 4208 | 4170 | 4246 | 150800 | 134100 | 167600 |
| 1979 | 7299 | 3569 | 12250 | 1742 | 1341 | 2142 | 41260 | 33810 | 48690 | 12020 | 10590 | 13440 | 10510 | 9415 | 11610 | 1942 | 1925 | 1959 | 74980 | 65820 | 84270 |
| 1980 | 17360 | 8521 | 29320 | 3903 | 3187 | 4616 | 10 | 80440 | 115700 | 56860 | 49800 | 63990 | 38660 | 34750 | 42590 | 5796 | 5744 | 5848 | 221300 | 198300 | 2500 |
| 1981 | 15620 | 7668 | 26300 | 7027 | 5475 | 8577 | 77070 | 63190 | 90930 | 24360 | 20360 | 28340 | 23210 | 20770 | 25630 | 5601 | 5551 | 5651 | 153400 | 135300 | 17180 |
| 1982 | 11590 | 5670 | 19530 | 3165 | 2516 | 3809 | 68410 | 56040 | 80660 | 41950 | 32780 | 51060 | 16740 | 14850 | 18620 | 6056 | 6002 | 6111 | 148300 | 130400 | 166100 |
| 1983 | 8353 | 4112 | 14070 | 3702 | 3018 | 4384 | 56110 | 46010 | 66190 | 31290 | 25740 | 36800 | 16480 | 14510 | 18470 | 2155 | 2136 | 2174 | 118400 | 105100 | 131800 |
| 1984 | 6028 | 2957 | 10110 | 3362 | 2447 | 4278 | 51910 | 49400 | 54400 | 29590 | 20800 | 38250 | 21470 | 18330 | 24610 | 3222 | 3193 | 3251 | 115700 | 105000 | 126400 |
| 1985 | 4715 | 2318 | 7958 | 2743 | 1914 | 3572 | 53690 | 50550 | 56840 | 36020 | 25280 | 46710 | 29700 | 25400 | 34000 | 5529 | 5479 | 5579 | 132600 | 119900 | 145100 |
| 1986 | 8162 | 4010 | 13690 | 3263 | 2381 | 4141 | 63870 | 60490 | 67260 | 57060 | 40690 | 73550 | 21380 | 18130 | 24630 | 6176 | 612 | 6231 | 160200 | 142100 | 178200 |
| 1987 | 11030 | 5423 | 18570 | 2351 | 1657 | 3047 | 60540 | 57410 | 63650 | 35870 | 25880 | 45910 | 13640 | 11610 | 1569 | 3081 | 305 | 310 | 12690 | 11410 | 13990 |
| 1988 | 6916 | 3377 | 11620 | 3430 | 2442 | 4422 | 66140 | 62370 | 69880 | 42680 | 31230 | 54070 | 11780 | 9924 | 13620 | 3286 | 3256 | 3316 | 134400 | 121300 | 147500 |
| 1989 | 6651 | 3256 | 11170 | 1685 | 1245 | 2128 | 59360 | 56310 | 62380 | 28170 | 20660 | 35700 | 14640 | 12390 | 16880 | 3197 | 3168 | 3226 | 113900 | 104500 | 123400 |
| 1990 | 3824 | 1875 | 6443 | 2691 | 2008 | 3370 | 58300 | 54840 | 61760 | 36870 | 26270 | 47600 | 11660 | 9885 | 13420 | 5051 | 5005 | 5096 | 118600 | 106800 | 130300 |
| 1991 | 1877 | 921 | 3148 | 2057 | 1566 | 2546 | 53790 | 50720 | 56880 | 35900 | 24820 | 46920 | 13040 | 11150 | 14940 | 2647 | 2623 | 2671 | 109400 | 97600 | 121100 |
| 1992 | 7534 | 3982 | 12740 | 8169 | 5447 | 10880 | 54100 | 50870 | 57350 | 37980 | 32120 | 43830 | 11970 | 10260 | 13680 | 2459 | 2437 | 248 | 12250 | 11390 | 131400 |
| 1993 | 9464 | 5911 | 1507 | 4359 | 3237 | 5494 | 760 | 030 | 3480 | 3220 | 110 | S290 | 8089 | 718 | 8990 | 2231 | 221 | 225 | 10950 | 8861 | 1304 |
| 1994 | 12950 | 8481 | 20330 | 4038 | 2903 | 5179 | 42440 | 40720 | 44150 | 30340 | 24070 | 36570 | 5167 | 4650 | 5683 | 1346 | 1334 | 1358 | 96700 | 8804 | 10620 |
| 1995 | 25550 | 18140 | 37410 | 3857 | 2582 | 5119 | 48970 | 47040 | 50900 | 39660 | 33760 | 45570 | 6830 | 6004 | 7655 | 1748 | 1732 | 1764 | 127000 | 116500 | 140200 |
| 1996 | 18830 | 13380 | 27640 | 5679 | 4053 | 7290 | 44620 | 42580 | 46680 | 29710 | 23280 | 36080 | 9200 | 8117 | 10290 | 2407 | 2385 | 2429 | 110800 | 101400 | 121700 |
| 1997 | 16190 | 11580 | 23740 | 6013 | 4248 | 7790 | 36740 | 35070 | 38410 | 24260 | 18490 | 30000 | 4575 | 4121 | 5032 | 1611 | 1597 | 1626 | 89840 | 81420 | 99180 |
| 199 | 8795 | 52 | 12510 | 6456 | 4513 | 8404 | 100 | 2520 | 680 | 16610 | 12780 | 20480 | 2604 | 2393 | 4 | 1526 | 1512 | 1540 | 100 | 58180 | 70090 |
| 1999 | 10510 | 6240 | 14970 | 6283 | 4354 | 8209 | 29560 | 27860 | 31260 | 16180 | 13040 | 19340 | 4192 | 3914 | 4472 | 1168 | 1157 | 1178 | 67900 | 61830 | 7405 |
| 2000 | 14360 | 8529 | 20440 | 6379 | 4521 | 8233 | 28390 | 26210 | 30580 | 17330 | 14050 | 20610 | 2378 | 2161 | 2595 | 533 | 528 | 538 | 69380 | 61930 | 7700 |
| 2001 | 15180 | 9012 | 21600 | 2505 | 1696 | 3305 | 29710 | 27470 | 31970 | 27500 | 23540 | 31460 | 4273 | 3912 | 4634 | 788 | 781 | 795 | 79930 | 72030 | 88040 |
| 2002 | 11070 | 6456 | 15880 | 2427 | 1594 | 3253 | 21320 | 19550 | 23090 | 14990 | 11690 | 17300 | 969 | 896 | 1042 | 504 | 500 | 509 | 50770 | 48860 | 56800 |
| 2003 | 9302 | 4852 | 13820 | 3381 | 2229 | 4539 | 33160 | 30710 | 35630 | 26550 | 21530 | 31590 | 3327 | 3009 | 3647 | 1192 | 1181 | 1203 | 76910 | 69460 | 84380 |
| 2004 | 11150 | 7552 | 14940 | 3319 | 2084 | 4555 | 28950 | 27010 | 30880 | 26390 | 20620 | 32120 | 2691 | 2465 | 2917 | 1283 | 1271 | 1295 | 73780 | 66530 | 81110 |
| 2005 | 13690 | 7934 | 19730 | 4410 | 2549 | 6291 | 27960 | 26210 | 29690 | 27010 | 21220 | 32750 | 1695 | 1542 | 1848 | 984 | 975 | 993 | 75770 | 67030 | 84690 |
| 2006 | 13760 | 8680 | 19130 | 5373 | 3555 | 7199 | 26170 | 24480 | 27860 | 23010 | 18170 | 27860 | 2545 | 2292 | 2797 | 1023 | 1014 | 1032 | 71900 | 64250 | 79690 |
| 2007 | 14300 | 8423 | 20420 | 4163 | 2631 | 5690 | 23920 | 22290 | 25540 | 23100 | 18950 | 27240 | 1389 | 1270 | 1509 | 954 | 945 | 963 | 67810 | 60080 | 75750 |
| 2008 | 17130 | 10360 | 24200 | 3870 | 2445 | 5311 | 28240 | 25870 | 30600 | 19090 | 14580 | 23570 | 3056 | 2728 | 3383 | 1764 | 1748 | 1780 | 73140 | 64300 | 82140 |
| 2009 | 25520 | 13430 | 37950 | 4618 | 2790 | 6452 | 27470 | 25640 | 29290 | 24030 | 19700 | 28380 | 2666 | 2423 | 2911 | 2069 | 2050 | 2088 | 86370 | 73150 | 99950 |
| 2010 | 8958 | 5270 | 12790 | 4669 | 3133 | 6198 | 29250 | 27330 | 31160 | 21500 | 17000 | 26050 | 2015 | 1838 | 2194 | 1078 | 1068 | 1088 | 67490 | 61020 | 74000 |
| 2011 | 28320 | 8011 | 48970 | 3663 | 2370 | 4949 | 37400 | 35110 | 39680 | 57760 | 42690 | 72780 | 4639 | 4183 | 5095 | 3045 | 3018 | 3072 | 134900 | 108300 | 161600 |
| 2012 | 22060 | 8110 | 36220 | 3329 | 2142 | 4501 | 27040 | 25250 | 28840 | 19260 | 15190 | 23350 | 1082 | 968 | 1195 | 881 | 873 | 889 | 73680 | 58860 | 88800 |

6.xv. Estimated 2SW salmon spawners for the six North American regions and North American total from the run reconstruction model.

| Spawner L | Labrador |  | Newfoundland |  |  |  | Quebec <br> Median | 5th perc. | 95th perc | Gulf Median | 5th perc. | Scotia-Fundy |  |  |  | USA | 5th perc. | North America |  |  | 95th perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Median | 5th perc. | 95th perc | Median | 5th perc. | 95th perc |  |  |  |  |  | 95th perc | Median | 5th perc. | 95th perc | Media |  | 95th perc N | Median | 5th perc. |  |
| 1970 | 9563 | 4407 | 16440 | 3236 | 2306 | 4167 | 28570 | 23420 | 33700 | 9978 | 8169 | 11780 | 6500 | 4695 | 8294 |  |  |  |  |  |  |
| 1971 | 13930 | 6582 | 23800 | 2982 | 2085 | 3876 | 14770 | 12110 | 17440 | 10440 | 8282 | 12560 | 7061 | 5598 | 8520 | 490 | 486 | 494 | 750 | 41020 | 60390 |
| 1972 | 12010 | - 5674 | 20430 | 3141 | 2208 | 4073 | 28970 | 23740 | 34170 | 200 | 22330 | 36050 | 10380 | - 8736 | 12030 | 1038 | 1029 | 1047 | 84990 | 73160 | 97440 |
| 1973 | 16290 | 785 | 28160 | 3842 | 2783 | 4891 | 29470 | 24150 | 4770 | 32210 | 25340 | 39080 | 6699 | 554 | 858 | 1100 | 90 | 110 | 89950 | 76320 | 1049 |
| 1974 | 16250 | 7562 | 27980 | 3140 | 2427 | -3847 | 35840 | 29380 | 42250 | 49060 | 38990 | 58990 | 14080 | 11940 | 16200 | 1147 | 113 | 1157 | 119900 | 103600 | 13700 |
| 1975 | 15560 | 7483 | 26500 | 4706 | 3451 | 5960 | 29780 | 24410 | 35130 | 28900 | 22670 | 35160 | 16360 | 13870 | 18850 | 1942 | 1925 | 1960 | 97550 | 84590 | 111700 |
| 1976 | 17490 | 8173 | 29970 | 3985 | 3001 | 4966 | 28300 | 23220 | 33400 | 24080 | 18330 | 29860 | 15510 | 12920 | 18100 | 1126 | 1116 | 1136 | 90760 | 77430 | 105800 |
| 1977 | 14970 | 6693 | 26110 | 2768 | 2182 | 3354 | 40780 | 33420 | 48070 | 51390 | 39990 | 62750 | 18850 | 15710 | 21950 | 643 | 637 | 649 | 129800 | 112300 | 147800 |
| 1978 | 11970 | 549 | 20680 | 3052 | 2467 | 5640 | 37330 | 30660 | 44080 | 15960 | 12060 | 19850 | 9419 | 7934 | 10890 | 3314 | 3284 | 3344 | 81330 | 70210 | 932 |
| 1979 | 6690 | 2960 | 11650 | 1616 | 1236 | 1999 | 16020 | 13130 | 18900 | 5767 | 4389 | 7154 | 6679 | 5650 | 7713 | 1509 | 1495 | 1523 | 38390 | 32930 | 442 |
| 1980 | 16470 | 7632 | 28430 | 3262 | 241 | 3887 | 4540 | 36480 | 2490 | 31510 | 24600 | 38380 | 21290 | 17690 | 24920 | 4263 | 4225 | 4301 | 121700 | 106200 | 138200 |
| 1981 | 15100 | 7148 | 25780 | - 6589 | 5104 | - 8078 | 2670 | 780 | 350 | 9735 | 5835 | 13640 | 10360 | 8249 | 12480 | 4334 | 295 | 4373 | 79050 | 67140 | 20 |
| 1982 | 10970 | 5049 | 18900 | 2769 | 2167 | 3369 | 33110 | 27150 | 39080 | 21170 | 12120 | 30290 | 7813 | 6185 | 9436 | 4643 | 4601 | 4685 | 80800 | 67160 | 94760 |
| 1983 | 7925 | 3684 | 13650 | 3279 | 2655 | 3907 | 21670 | 17740 | 25530 | 13960 | 8497 | 19420 | 4210 | 2662 | 5744 | 1769 | 1753 | 1785 | 53040 | 44090 | 62220 |
| 1984 | 5518 | 2447 | 9596 | 3179 | 2286 | 4068 | 27090 | 24910 | 29260 | 26030 | 17330 | 34750 | 17510 | 14540 | 20460 | 2547 | 2524 | 2570 | 82060 | 71540 | 92620 |
| 1985 | 442 | 202 | 7664 | 2728 | 1906 | -3557 | 25870 | 23040 | 28710 | 35030 | 24300 | 45750 | 24640 | 20490 | 28760 | 4884 | 4840 | 4928 | 97710 | 85290 | 11020 |
| 1986 | 7695 | 3543 | 13230 | 3229 | 2350 | 03 | 29650 | 26720 | 32620 | 55410 | 38920 | 71780 | 18440 | 15290 | 21600 | 5570 | 5520 | 5620 | 120300 | 102300 | 138200 |
| 1987 | 10390 | 4790 | 17940 | 2335 | 1642 | - 3028 | 26320 | 23790 | 28840 | 3870 | 23910 | 43800 | 12200 | 10 | 14190 | 278 | 275 | 280 | 8824 | 75650 | 101100 |
| 1988 | 6206 | 266 | 10910 | 3412 | 2426 | 4393 | 31510 | 28170 | 34850 | 41310 | 29960 | 52650 | 10320 | 8525 | 12130 | 3038 | 3011 | 3065 | 96000 | 83040 | 108900 |
| 1989 | 190 | 95 | 10710 | 1681 | 1241 | 19 | 30040 | 27350 | 32720 | 26940 | 19450 | 34450 | 14310 | 12060 | 16540 | 2800 | 2775 | 2825 | 82150 | 72920 | 91550 |
| 1990 | 3467 | 1518 | 6086 | 2669 | 1991 | 3346 | 29870 | 26680 | 33040 | 35770 | 25130 | 46410 | 11000 | 9249 | 12750 | 4356 | 4317 | 4395 | 87270 | 75620 | 98920 |
| 1991 | 1784 | 828 | 3055 | 2045 | 1557 | 2534 | 24140 | 21450 | 26820 | 35070 | 23980 | 46090 | 11660 | 9824 | 13490 | 2416 | 2394 | 2438 | 77190 | 65510 | 88790 |
| 1992 | 6752 | 3200 | 11950 | 8117 | 5413 | 10830 | 23620 | 20790 | 26450 | 36800 | 31000 | 42620 | 10820 | 9165 | 12470 | 2292 | 2271 | 2313 | 88650 | 80260 | 97390 |
| 1993 | 9077 | - 5524 | 14690 | 4315 | - 3191 | 5438 | 18220 | 16890 | 40 | 260 | 290 | 2750 | 6931 | - 6049 | 811 | 2065 | 2046 | 2084 | 83670 | 2730 | 104500 |
| 199 | 12460 | 991 | 19840 | 386 | 2771 | 500 | 17860 | 16570 | 19140 | 29630 | 23420 | 35840 | 4390 | 3899 | 488 | 1344 | 1332 | 1356 | 70010 | 61460 | 79490 |
| 1995 | 25090 | 17680 | 36950 | 3706 | 2460 | 4956 | 25270 | 23850 | 26670 | 39130 | 33220 | 44990 | 6465 | 5645 | 7287 | 1748 | 1732 | 1764 | 101700 | 91340 | 114800 |
| 1996 | 18450 | 13000 | 27260 | 5498 | 3911 | 7091 | 21930 | 20320 | 23540 | 28850 | 22520 | 35160 | 8381 | 7320 | 9449 | 2407 | 2385 | 2429 | 85900 | 76730 | 96610 |
| 1997 | 15980 | 11370 | 23530 | 5872 | 4136 | 7622 | 18120 | 16820 | 19440 | 23440 | 17740 | 29130 | 3975 | 3535 | 4414 | 1611 | 1597 | 1626 | 69420 | 61110 | 78740 |
| 1998 | 8592 | 5015 | 12290 | 6352 | 4425 | 8272 | 16820 | 15500 | 18120 | 16120 | 12340 | 19920 | 2273 | 2073 | 2474 | 1526 | 1512 | 1540 | 51680 | 45830 | 57610 |
| 1999 | 10240 | 5972 | 14690 | 6206 | 4291 | 8122 | 20370 | 18800 | 21960 | 15390 | 12270 | 18490 | 3732 | 3457 | 4006 | 1168 | 1157 | 1178 | 57110 | 51110 | 63180 |
| 2000 | 14100 | 8264 | 20160 | 6221 | 4387 | 8048 | 19500 | 17410 | 21600 | 16730 | 13500 | 19970 | 2178 | 1966 | 2392 | 1587 | 1573 | 1601 | 60330 | 52900 | 67930 |
| 2001 | 14860 | 8696 | 21270 | 2432 | 1645 | 3224 | 20050 | 18160 | 21960 | 26570 | 22660 | 30510 | 4009 | 3656 | 4360 | 1491 | 1478 | 1504 | 69430 | 61650 | 77400 |
| 2002 | 10870 | 6256 | 15670 | 2381 | 1562 | 3200 | 15120 | 13460 | 16800 | 13950 | 11170 | 16750 | 786 | 718 | 853 | 511 | 506 | 516 | 43630 | 37750 | 49590 |
| 2003 | 9069 | 4620 | 13580 | 3307 | 2170 | 4452 | 24650 | 22260 | 27050 | 25730 | 20770 | 30690 | 3120 | 2803 | 3435 | 1192 | 1181 | 1203 | 67070 | 59700 | 742 |
| 2004 | 10880 | 7288 | 14670 | 3255 | 2032 | 4466 | 20540 | 18710 | 22390 | 25540 | 19840 | 31170 | 2575 | 2353 | 2795 | 1283 | 1271 | 1295 | 64060 | 56900 | 71280 |
| 2005 | 13410 | 7659 | 19450 | 4325 | 2479 | 6172 | 20500 | 18850 | 22170 | 26050 | 20350 | 31750 | 1588 | 1440 | 1736 | 1088 | 1078 | 1098 | 67000 | 58260 | 75820 |
| 2006 | 13540 | 8461 | 18900 | 5290 | 3485 | 7099 | 19030 | 17460 | 20600 | 22190 | 17370 | 27010 | 2394 | 2147 | 2639 | 1419 | 1406 | 1432 | 63880 | 56250 | 71640 |
| 2007 | 14070 | 8189 | 20180 | 4097 | 2593 | 5622 | 17220 | 15670 | 18750 | 22240 | 18120 | 26330 | 1280 | 1165 | 1395 | 1189 | 1178 | 1200 | 60080 | 52380 | 67940 |
| 2008 | 16900 | 10130 | 23960 | 3772 | 2368 | 5188 | 21770 | 19450 | 24070 | 18240 | 13830 | 22690 | 2958 | 2636 | 3282 | 2809 | 2784 | 2834 | 66440 | 57700 | 7539 |
| 2009 | 25300 | 13210 | 37720 | 4559 | 2748 | 6371 | 20960 | 19190 | 22730 | 23140 | 18850 | 27410 | 2548 | 2310 | 2784 | 2292 | 2271 | 2313 | 78820 | 65570 | 92320 |
| 2010 | 8760 | 5072 | 12580 | 4570 | 3047 | 6081 | 23380 | 21520 | 25250 | 20670 | 16210 | 25150 | 1883 | 1708 | 2058 | 1482 | 1469 | 1495 | 60770 | 54340 | 67220 |
| 2011 | 28180 | 7873 | 48830 | 3625 | 2342 | 4904 | 29390 | 27170 | 31600 | 56270 | 41400 | 71200 | 4557 | 4103 | 5013 | 3872 | 3837 | 3907 | 125900 | 99560 | 152700 |
| 2012 | 21970 | 8017 | 36120 | 3304 | 2127 | 4479 | 20740 | 19020 | 22460 | 18570 | 14510 | 22610 | 1030 | 917 | 1143 | 2056 | 2038 |  | 67640 | 52870 | 82680 |

6.xvi. North American pre-fishery abundance (PFA) estimates from the run reconstruction model.

|  | PFA 1SW non-maturing |  |  | PFA 1SW maturing |  | PFA total (maturing and non-maturing) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Median | 5th perc. | 95th perc | Median | 5th perc. | 95th perc | Median | 5th perc. | 95th perc |
| 1971 | 713700 | 650100 | 778700 | - 520000 | 484700 | 560500 | 1235000 | 1165000 | 1306000 |
| 1972 | 740400 | 684800 | 801400 | - 520700 | 491300 | 553700 | 1262000 | 1203000 | 1325000 |
| 1973 | 901600 | 820700 | 986000 | - 666900 | 635800 | 698800 | 1569000 | 1487000 | 1654000 |
| 1974 | 811900 | 750700 | 877700 | - 699100 | 662100 | 739300 | 1511000 | 1446000 | 1582000 |
| 1975 | 904800 | 839300 | 974600 | - 798500 | 746300 | 861100 | 1705000 | 1627000 | 1790000 |
| 1976 | 835500 | 765700 | 910500 | - 798400 | 751500 | 849900 | 1635000 | 1556000 | 1720000 |
| 1977 | 667100 | 606900 | 729800 | - 636300 | 594900 | 682100 | 1304000 | 1236000 | 1376000 |
| 1978 | 396700 | 368200 | 426800 | - 410700 | 383000 | 439400 | 807500 | 770800 | 846100 |
| 1979 | 837200 | 772100 | 908700 | - 589500 | 557400 | 623400 | 1427000 | 1356000 | 1504000 |
| 1980 | 711400 | 655100 | 771500 | 832400 | 781000 | 892700 | 1545000 | 1475000 | 1621000 |
| 1981 | 667000 | 620900 | 716000 | 911400 | 849000 | 981700 | 1579000 | 1506000 | 1658000 |
| 1982 | 560400 | 523600 | 599900 | 765800 | 715300 | 820300 | 1327000 | 1267000 | 1390000 |
| 1983 | 341900 | 311900 | 374700 | 511400 | 479900 | 545400 | 853700 | 812600 | 897700 |
| 1984 | 360500 | 328800 | 395100 | 539700 | 506100 | 573900 | 900400 | 854800 | 948100 |
| 1985 | 535800 | 492300 | 582700 | 658700 | 617000 | 701300 | 1194000 | 1135000 | 1257000 |
| 1986 | 569100 | 521100 | 619500 | 835800 | 779300 | 893900 | 1405000 | 1334000 | 1479000 |
| 1987 | 518900 | 481400 | 558800 | 801100 | 749600 | 858000 | 1320000 | 1261000 | 1385000 |
| 1988 | 422300 | 389500 | 457100 | 850000 | 789500 | 912700 | 1273000 | 1205000 | 1343000 |
| 1989 | 334100 | 305300 | 365100 | 595000 | 557300 | 635300 | 929600 | 883500 | 978100 |
| 1990 | 298500 | 273300 | 326000 | 562400 | 527200 | 597900 | 861100 | 817300 | 906100 |
| 1991 | 330800 | 308100 | 355600 | 415400 | 390300 | 441100 | 746300 | 712400 | 782200 |
| 1992 | 217000 | 184900 | 252400 | 577700 | 531700 | 624200 | 795100 | 737000 | 854600 |
| 1993 | 157300 | 139800 | 177200 | 545500 | 483300 | 608200 | 703300 | 637300 | 769900 |
| 1994 | 193400 | 171600 | 219200 | 329600 | 301500 | 358600 | 523500 | 486400 | 563000 |
| 1995 | 190600 | 171500 | 212300 | 382200 | 345300 | 419700 | 573100 | 530500 | 617200 |
| 1996 | 161000 | 144800 | 179400 | 555700 | 500600 | 614200 | 716900 | 658700 | 778700 |
| 1997 | 111700 | 100500 | 124000 | 363100 | 330900 | 403900 | 475200 | 440000 | 518000 |
| 1998 | 102100 | 90550 | 115000 | 442800 | 390900 | 495300 | 545200 | 491000 | 600000 |
| 1999 | 107400 | 94180 | 122000 | 443100 | 390800 | 495300 | 550500 | 495800 | 605600 |
| 2000 | 121900 | 107500 | 138100 | 525900 | 464000 | 587700 | 648100 | 583200 | 713100 |
| 2001 | 84720 | 74560 | 95920 | 386500 | 337800 | 435600 | 471400 | 420800 | 522600 |
| 2002 | 114600 | 100900 | 129600 | 389100 | 347700 | 430600 | 503900 | 459200 | 549200 |
| 2003 | 112300 | 98940 | 127100 | 421400 | 384400 | 458400 | 533900 | 493200 | 575000 |
| 2004 | 115500 | 100500 | 132100 | 448200 | 415000 | 481500 | 564000 | 526300 | 601900 |
| 2005 | 110400 | 96920 | 125300 | 546700 | 471100 | 622400 | 657400 | 579700 | 735100 |
| 2006 | 105100 | 91740 | 119800 | 548500 | 470000 | 627000 | 653600 | 573300 | 734500 |
| 2007 | 116200 | 101400 | 133000 | 474300 | 408800 | 539400 | 590900 | 522500 | 659100 |
| 2008 | 135400 | 114400 | 158700 | 596100 | 530900 | 661200 | 731800 | 661900 | 802100 |
| 2009 | 108200 | 96070 | 121500 | 381900 | 323800 | 440100 | 490200 | 430500 | 550500 |
| 2010 | 205800 | 166200 | 248200 | 504900 | 466400 | 543300 | 711000 | 653500 | 769800 |
| 2011 | 116800 | 94540 | 141100 | 671700 | 489900 | 854900 | 788800 | 605000 | 974000 |

## Annex 7: Glossary of acronyms used in this Report

1SW (One-Sea-Winter). Maiden adult salmon that has spent one winter at sea.
2SW (Two-Sea-Winter). Maiden adult salmon that has spent two winters at sea.
ACOM (Advisory Committee) of ICES. The Committee works on the basis of scientific assessment prepared in the ICES expert groups. The advisory process includes peer review of the assessment before it can be used as the basis for advice. The Advisory Committee has one member from each member country under the direction of an independent chair appointed by the Council.

BCI (Bayesian Credible Interval). The Bayesian equivalent of a confidence interval. If the $90 \% \mathrm{BCI}$ for a parameter A is 10 to 20 , there is a $90 \%$ probability that A falls between 10 and 20.

BHSRA (Bayesian Hierarchical Stock and Recruitment Approach). Models for the analysis of a group of related stock-recruit datasets. Hierarchical modelling is a statistical technique that allows the modelling of the dependence among parameters that are related or connected through the use of a hierarchical model structure. Hierarchical models can be used to combine data from several independent sources.
$\mathbf{C \& R}$ (Catch and Release). Catch and release is a practice within recreational fishing intended as a technique of conservation. After capture, the fish are unhooked and returned to the water before experiencing serious exhaustion or injury. Using barbless hooks, it is often possible to release the fish without removing it from the water (a slack line is frequently sufficient).
CL, i.e. Slim (Conservation Limit). Demarcation of undesirable stock levels or levels of fishing activity; the ultimate objective when managing stocks and regulating fisheries will be to ensure that there is a high probability that undesirable levels are avoided.

COSEWIC (Committee on the Status of Endangered Wildlife in Canada). COSEWIC is the organization that assesses the status of wild species, subspecies, varieties, or other important units of biological diversity, considered to be at risk of extinction in Canada. COSEWIC uses scientific, Aboriginal traditional and community knowledge provided by experts from governments, academia and other organizations. Summaries of assessments on Atlantic salmon are currently available to the public on the COSEWIC website (www.cosewic.gc.ca)

Cpue (Catch Per Unit of Effort). A derived quantity obtained from the independent values of catch and effort.

CWT (Coded Wire Tag). The CWT is a length of magnetized stainless steel wire 0.25 mm in diameter. The tag is marked with rows of numbers denoting specific batch or individual codes. Tags are cut from rolls of wire by an injector that hypodermically implants them into suitable tissue. The standard length of a tag is 1.1 mm .

DFO (Department of Fisheries and Oceans). DFO and its Special Operating Agency, the Canadian Coast Guard, deliver programs and services that support sustainable use and development of Canada's waterways and aquatic resources.

DNA (Deoxyribonucleic Acid). DNA is a nucleic acid that contains the genetic instructions used in the development and functioning of all known living organisms (with the exception of RNA- Ribonucleic Acid viruses). The main role of DNA molecules is the long-term storage of information. DNA is often compared to a set of blueprints,
like a recipe or a code, since it contains the instructions needed to construct other components of cells, such as proteins and RNA molecules.

DST (Data Storage Tag). A miniature data logger with sensors including salinity, temperature, and depth that is attached to fish and other marine animals.

ECOKNOWS (Effective use of Ecosystems and biological Knowledge in fisheries). The general aim of the ECOKNOWS project is to improve knowledge in fisheries science and management. The lack of appropriate calculus methods and fear of statistical over partitioning in calculations, because of the many biological and environmental influences on stocks, has limited reality in fisheries models. This reduces the biological credibility perceived by many stakeholders. ECOKNOWS will solve this technical estimation problem by using an up-to-date methodology that supports more effective use of data. The models will include important knowledge of biological processes.

ENPI CBC (European Neighbourhood and Partnership Instrument Cross-Border Cooperation). ENPI CBC is one of the financing instruments of the European Union. The ENPI programmes are being implemented on the external borders of the EU. It is designed to target sustainable development and approximation to EU policies and standards; supporting the agreed priorities in the European Neighbourhood Policy Action Plans, as well as the Strategic Partnership with Russia.

FWI (Framework of Indicators). The FWI is a tool used to indicate if any significant change in the status of stocks used to inform the previously provided multi-annual management advice has occurred.

GRAASP (Genetically based Regional Assignment of Atlantic Salmon Protocol). GRAASP was developed and validated by twelve European genetic research laboratories. Existing and new genetic data were calibrated and integrated in a purpose built electronic database to create the assignment baseline. The unique database created initially encompassed 32002 individuals from 588 rivers. The baseline data, based on a suite of 14 microsatellite loci, were used to identify the natural evolutionary regional stock groupings for assignment.

ICPR (The International Commission for the Protection of the River Rhine). ICPR coordinates the ecological rehabilitation programme involving all countries bordering the river Rhine. This programme was initiated in response to catastrophic river pollution in Switzerland in 1986 which killed hundreds of thousands of fish. The programme aims to bring about significant ecological improvement of the Rhine and its tributaries enabling the re-establishment of migratory fish species such as salmon.

ISAV (Infectious Salmon Anemia Virus). ISAV is a highly infectious disease of Atlantic salmon caused by an enveloped virus.

LE (Lagged Eggs). The summation of lagged eggs from 1 and 2 sea winter fish is used for the first calculation of PFA.

LMN (Labrador Métis Nation). LMN is one of four subsistence fisheries harvesting salmonids in Labrador. LMN members are fishing in southern Labrador from Fish Cove Point to Cape St Charles.

MSY (Maximum Sustainable Yield). The largest average annual catch that may be taken from a stock continuously without affecting the catch of future years; a constant longterm MSY is not a reality in most fisheries, where stock sizes vary with the strength of year classes moving through the fishery.

MSW (Multi-Sea-Winter). A MSW salmon is an adult salmon which has spent two or more winters at sea and may be a repeat spawner.

NG (Nunatsiavut Government). NG is one of four subsistence fisheries harvesting salmonids in Labrador. NG members are fishing in the northern Labrador communities.

NSERC (Natural Sciences and Engineering Research Council of Canada). NSERC is a Canadian government agency that provides grants for research in the natural sciences and in engineering. Its mandate is to promote and assist research. Council supports a project to develop a standardized genetic database for North America.

OSPAR is the mechanism by which fifteen Governments of the west coasts and catchments of Europe, together with the European Community, cooperate to protect the marine environment of the Northeast Atlantic. It started in 1972 with the Oslo Convention against dumping. It was broadened to cover land-based sources and the offshore industry by the Paris Convention of 1974. These two conventions were unified, updated and extended by the 1992 OSPAR Convention. The new annex on biodiversity and ecosystems was adopted in 1998 to cover non-polluting human activities that can adversely affect the sea.

PFA (Pre-Fishery Abundance). The numbers of salmon estimated to be alive in the ocean from a particular stock at a specified time. In the previous version of the stock complex Bayesian PFA forecast model two productivity parameters are calculated, for the maturing (PFAm) and non-maturing (PFAnm) components of the PFA. In the updated version only one productivity parameter is calculated, and used to calculate total PFA, which is then split into PFAm and PFAnm based upon the proportion of PFAm (p.PFAm).

PGA (The Probabilistic-based Genetic Assignment model). An approach to partition the harvest of mixed-stock fisheries into their finer origin parts. PGA uses Monte Carlo sampling to partition the reported and unreported catch estimates to continent, country and within country levels.

PGCCDBS The Planning Group on Commercial Catches, Discards and Biological Sampling.
PGNAPES (Planning Group on Northeast Atlantic Pelagic Ecosystem Surveys). PGNAPES coordinates international pelagic surveys in the Norwegian Sea and to the West of the British Isles, directed in particular towards Norwegian Spring-spawning Herring and Blue Whiting. In addition, these surveys collect environmental information. The work in the group has progressed as planned.

PIT (Passive Integrated Transponder). PIT tags use radio frequency identification technology. PIT tags lack an internal power source. They are energized on encountering an electromagnetic field emitted from a transceiver. The tag's unique identity code is programmed into the microchip's nonvolatile memory

PSAT (Pop-up Satellite Archival Tags). Used to track movements of large, migratory, marine animals. A PSAT is an archival tag (or data logger) that is equipped with a means to transmit the data via satellite.

PSU (Practical Salinity Units). PSU are used to describe salinity: a salinity of $35 \%$ equals 35 PSU.

Q Areas for which the Ministère des Ressources naturelles et de la Faune manages the salmon fisheries in Québec.

RR model (Run-Reconstruction model). RR model is used to estimate PFA and national CLs.

RVS (Red Vent Syndrome). This condition has been noted since 2005, and has been linked to the presence of a nematode worm, Anisakis simplex. This is a common parasite of marine fish and is also found in migratory species. The larval nematode stages in fish are usually found spirally coiled on the mesenteries, internal organs and less frequently in the somatic muscle of host fish.

SALSEA (Salmon at Sea). SALSEA is an international programme of co-operative research designed to improve understanding of the migration and distribution of salmon at sea in relation to feeding opportunities and predation. It differentiates between tasks which can be achieved through enhanced coordination of existing ongoing research, and those involving new research for which funding is required.

SARA (Species At Risk Act). SARA is a piece of Canadian federal legislation which became law in Canada on December 12, 2002. It is designed to meet one of Canada's key commitments under the International Convention on Biological Diversity. The goal of the Act is to protect endangered or threatened organisms and their habitats. It also manages species which are not yet threatened, but whose existence or habitat is in jeopardy. SARA defines a method to determine the steps that need to be taken in order to help protect existing relatively healthy environments, as well as recover threatened habitats. It identifies ways in which governments, organizations, and individuals can work together to preserve species at risk and establishes penalties for failure to obey the law.

SCICOM (Science Committee) of ICES. SCICOM is authorized to communicate to third-parties on behalf of the Council on science strategic matters and is free to institute structures and processes to ensure that inter alia science programmes, regional considerations, science disciplines, and publications are appropriately considered.

SER (Spawning Escapement Reserve). The CL increased to take account of natural mortality between the recruitment date (assumed to be 1st January) and the date of return to homewaters.

SFA (Salmon Fishing Areas). Areas for which the Department of Fisheries and Oceans (DFO) Canada manages the salmon fisheries.

SGBICEPS (The Study Group on the Identification of Biological Characteristics For Use As Predictors Of Salmon Abundance). The ICES study group established to complete a review of the available information on the life-history strategies of salmon and changes in the biological characteristics of the fish in relation to key environmental variables.

SGBYSAL (Study Group on the Bycatch of Salmon in Pelagic Trawl Fisheries). The ICES study group that was established in 2005 to study Atlantic salmon distribution at sea and fisheries for other species with a potential to intercept salmon.

SGEFISSA (Study Group on Establishing a Framework of Indicators of Salmon Stock Abundance). SGEFISSA is a study group established by ICES and met in November 2006.

SGERAAS (Study Group on Effectiveness of Recovery Actions for Atlantic Salmon). SGERAAS is the previous acronym for WGERAAS (Working Group on Effectiveness of Recovery Actions for Atlantic Salmon).

SGSSAFE (Study Group on Salmon Stock Assessment and Forecasting). The study group established to work on the development of new and alternative models for forecasting Atlantic salmon abundance and for the provision of catch advice.

Slim, i.e. CL (Conservation Limit). Demarcation of undesirable stock levels or levels of fishing activity; the ultimate objective when managing stocks and regulating fisheries will be to ensure that there is a high probability that the undesirable levels are avoided.

SSGEF (SCICOM Steering Group on Understanding Ecosystem Functioning). SSGEF is one of five Steering Groups of SCICOM (Science Committee of ICES). Chair: Graham Pierce (UK); term of office: January 2012-December 2014.

SST (Sea surface temperatures). SST is the water temperatures close to the surface. In practical terms, the exact meaning of surface varies according to the measurement method used. A satellite infrared radiometer indirectly measures the temperature of a very thin layer of about 10 micrometres thick of the ocean which leads to the phrase skin temperature. A microwave instrument measures subskin temperature at about 1 mm . A thermometer attached to a moored or drifting buoy in the ocean would measure the temperature at a specific depth, (e.g. at one meter below the sea surface). The measurements routinely made from ships are often from the engine water intakes and may be at various depths in the upper 20 m of the ocean. In fact, this temperature is often called sea surface temperature, or foundation temperature.

SVC (Spring Viraemia of Carp). SVC is a contagious and potentially fatal viral disease affecting fish. As its name implies, SVC may be seen in carp in spring. However, SVC may also be seen in other seasons (especially in autumn) and in other fish species including goldfish and the European wells catfish. Until recently, SVC had only been reported in Europe and the Middle East. The first cases of SVC reported in the United States were in spring 2002 in cultivated ornamental common carp (Koi) and wild common carp. The number of North American fish species susceptible to SVC is not yet known.

TAC (Total Allowable Catch). TAC is the quantity of fish that can be taken from each stock each year.

WFD (Water Framework Directive). Directive 2000/60/EC (WFD) aims to protect and enhance the water environment, updates all existing relevant European legislation, and promotes a new approach to water management through river-based planning. The Directive requires the development of River Basin Management Plans (RBMP) and Programmes of Measures (PoM) with the aim of achieving Good Ecological Status or, for artificial or more modified waters, Good Ecological Potential

WGBAST (Baltic Salmon and Trout Assessment Working Group). WGBAST took place in Uppsala, Sweden, 15-23 March 2012, chaired by Johan Dannewitz (Sweden). Main tasks of group are: address generic ToRs for Fish Stock Assessment Working Groups; evaluate estimates of salmon misreporting by Poland based on new data from Poland, from the EC inspections, logbooks, VMS and other relevant data sources; evaluate the possible reasons for the low at-sea survival of salmon stocks, including new information from the 2011 Salmon Summit; prepare for a benchmark assessment of the salmon stocks in autumn 2012 and others.

WGERAAS (Working Group on Effectiveness of Recovery Actions for Atlantic Salmon). WGERAAS had been established by ICES. The task of the study group is to provide a review of examples of successes and failures in wild salmon restoration and rehabilitation and develop a classification of activities which could be recommended under
various conditions or threats to the persistence of populations. The Working Group has had its first meeting in Belfast in February 2013. The next meeting is scheduled for February 2014 at ICES in Copenhagen.

WGF (West Greenland Fishery). Regulatory measures for the WGF have been agreed by the West Greenland Commission of NASCO for most years since NASCO's establishment. These have resulted in greatly reduced allowable catches in the WGF, reflecting declining abundance of the salmon stocks in the area.

WGRECORDS (Working Group on the Science Requirements to Support Conservation, Restoration and Management of Diadromous Species). WGRECORS was reconstituted as a Working Group from the Transition Group on the Science Requirements to Support Conservation, Restoration and Management of Diadromous Species (TGRECORDS).

WKADS (Workshop on Age Determination of Salmon). WKADS took place in Galway, Ireland, January 18th to 20th 2011, with the objectives of reviewing, assessing, documenting and making recommendations on current methods of ageing Atlantic salmon. The Workshop focused primarily on digital scale reading to measure age and growth with a view to standardization.

WKADS2 (A second Workshop on Age Determination of Salmon). Took place from September 4th to 6th, 2012 in Derry ~ Londonderry, Northern Ireland to addressed recommendations made at the previous WKADS meeting (2011) (ICES CM 2011/ACOM:44) to review, assess, document and make recommendations for ageing and growth estimations of Atlantic salmon using digital scale reading, with a view to standardization. Available tools for measurement, quality control and implementation of inter-laboratory QC were considered.

WKDUHSTI (Workshop on the Development and Use of Historical Salmon Tagging Information from Oceanic Areas). This workshop, established by ICES, was held in February 2007.

WKSHINI (Workshop on Salmon historical information-new investigations from old tagging data). This workshop met from 18-20 September 2008 in Halifax, Canada.

WKLUSTRE (Workshop on Learning from Salmon Tagging Records). This ICES Workshop established to complete compilation of available data and analyses of the resulting distributions of salmon at sea.

This glossary has been extracted from various sources, but chiefly the EU SALMODEL report (Crozier et al., 2003).

## Annex 8: NASCO has requested ICES to identify relevant data deficiencies, monitoring needs and research requirements

The Working Group recommends that it should meet in 2014 to address questions posed by ICES, including those posed by NASCO. The Working Group intends to convene in the headquarters of ICES in Copenhagen, Denmark from 18 to 27 March 2014.

## List of recommendations

1 ) The Working Group recommends that further work be undertaken to address the issues raised by the second Workshop on Age Determination of Salmon (WKADS 2). The following issues were identified and the Working Group recommended that these should be followed up:
1.1) An inter-lab calibration exercise should be held remotely in the next two to four years.
1.2 ) Reference scale images and accompanying details should be hosted on ICES age readers forum website.
1.3 ) The importance of the initial positioning of the line on a scale along which measurement are made, should be emphasized to all readers.
2 ) The Working Group recommended the establishment of a Regional Coordination Group (RCG) for diadromous species to consider the unified collection of data on all salmon stocks (as well as eel).
3 ) The Working Group recommended that an Atlantic salmon stock annex should be developed using an agreed template and country specific inputs should be prepared for the 2014 meeting with a view to finalizing the document at that time.

4 ) The Working Group recommends that the IASRB support the further development of the project outlined by the NASCO Sub-Group on the Future Direction of Research on Marine Survival of Salmon.

5 ) The Working Group welcomed the opportunistic assessment of the incidence of salmon bycatch in pelagic fisheries at Iceland and recommends that similar sampling should continue in order to provide further information on the bycatch of salmon in pelagic fisheries in this area.
6 ) The Working Group recommends that consideration be given to further investigations involving ultrasonic tagging and release of non-maturing salmon captured at Greenland.

7 ) The Working Group recommends that sampling of the Labrador and SaintPierre et Miquelon fisheries be continued and expanded (i.e. sample size, geographic coverage, tissue samples, seasonal distribution of the samples) in future years and analysed using the North American genetic baseline to improve the information on biological characteristics and stock origin of salmon harvested in these mixed-stock fisheries.
8 ) The Working Group recommends that additional data collection be considered in Labrador to better estimate salmon returns in that region.
9 ) The Working Group supports the efforts of the Geenlandic authorities to improve catch data collection and recommends that the authorities facili-
tate the coordination of sampling within factories receiving Atlantic salmon, if landings at factories are allowed in 2013.
10 ) The Working Group recommends that the Greenland catch reporting system continues and that logbooks be provided to all fishers. Efforts should continue to encourage compliance with the voluntary logbook system. Detailed statistics related to catch and effort should be made available to the Working Group for assessment.
11 ) The Working Group recommends that arrangements be made to enable sampling in Nuuk as a significant amount of salmon is landed in this community on an annual basis.
12 ) The Working Group recommends that the longer time-series of sampling data from West Greenland should be analysed to assess the extent of the variation in condition over the time period corresponding to the large variation in productivity as identified by the NAC and NEAC assessment and forecast models.

13 ) The Working Group recommends a continuation and expansion of the broad geographic sampling programme (multiple NAFO divisions) to more accurately estimate continent of origin in the Greenland mixed-stock fishery.

# Annex 9: Response of WGNAS 2013 to Technical Minutes of the Review Group (ICES 2012a) 

As per the request of the ICES Review Group (RG) this section is the response of the Working Group on North Atlantic Salmon (WGNAS) to the Technical Minutes of the RG provided in Annex 10 of ICES (2012a). The points are addressed in the same order as they were listed in the Technical Minutes. This section also provides a response to some additional comments and questions from the Chair of the RG, which were received by WGNAS prior to its 2013 meeting:

## Report structure

The RG commented that it would help reviewers as well as communication with the general public if a stock annex was provided, detailing the methodology used to conduct stock assessment and to provide catch advice (similar to how it is done in other ICES assessment WG reports).

The Working Group considered this request informed by the progress made by the Baltic Salmon Group in developing such an annex as part of their recent InterBenchmark Protocol exercise (ICES 2012b) and further examples provided by ICES. The Working Group agreed to take forward the development of a specific Atlantic salmon stock annex, but had insufficient time to complete the task during the 2013 meeting. Initial progress was made in completing sections of a draft, by compiling information contained in earlier WGNAS reports and from other sources. The Working Group recommended that the provision of country-specific inputs would be best addressed by identifying this as a specific requirement for the 2014 meeting, using an agreed template. This will be developed by correspondence over the year in advance of the 2014 meeting.

## Section 3: Northeast Atlantic Commission

| RG Comment | WGNAS RESPONSE |
| :---: | :---: |
| Two assumptions used in the NEAC model were a bit surprising: (1) the use of a constant mortality rate of 0.03 per month for non-mature 1SW, and (2) constant maturation rate for 1 SW of $78 \%$ (Table 3.6.4.5). The percentage of 1SW in the reported catch for the northern and southern NEAC countries varied among years and reached the lowest value in 2011 for both regions (Figures 3.1.6.1 and 3.1.6.2), which suggests that mortality rates of immature 1SW fish or maturation rate (or both) changed during the time period, possibly in monotonic fashion. The effects of violating these assumptions should be evaluated. | 1. Mortality rate: The natural mortality rate for salmon after they recruit to the distant water fisheries has been the subject of much discussion. The Working Group originally used a value of 0.01 , but this was modified to 0.03 following a detailed review as part of the EU SALMODEL project (Crozier et al., 2003; ICES, 2002). While mortality may be expected to vary among years and may also be different for maturing and nonmaturing 1SW recruits, the WG has not had data on which to base the use of different values, or values that change over time. However, this is now being further investigated within the EU ECOKNOWS project and Bayesian modelling may provide alternative approaches in future. <br> 2. Maturation rate: Maturation rate is handled differently in different models. <br> Run-reconstruction model: Maturation rate is not an input to the model. <br> Forecast model: The proportion maturing and |

RG COMMENT WGNAS RESPONSE
the mortality rate in the second year at sea are confounded parameters in the model and a strong prior, such as mortality rate known or proportion maturing known, must be set to resolve the other. Some consequences on model inferences are discussed in Section 2 in this year's report.

Catch option model: The maturation rate is applied only to the 1 SW catch to determine the (very small) numbers that remain at sea for another year. The value used (0.78) is based upon analysis of vitelogenin in blood samples collected from salmon caught in the Faroes fishery in about 1984. No other estimates have been obtained (and no fishery has operated for at least 10 years), and so no information is available on variation between years. The WG has therefore agreed to use a value of $0.78+/-0.1$

The pseudo-stock-recruitment relationship (hockey-stick) between Pre-fishery Abundance (PFA) and lagged-egg (used for the derivation of provisional national conservation limits) is assumed to be static over time. Is this a reasonable assumption given the observed declining trends in marine survival? It was also unclear from the WG report whether or not the lagged-egg deposition accounted for in-river mortality associated with catch-and-release. A table presenting the model parameters would be useful.

It seems surprising that retention fisheries still occurred in some countries despite the fact that returns and spawner abundance were lower than the conservation limits set by these countries (e.g. Figures 3.3.3.1b and 3.3.3.1h). Presumably this is because the target escapement was met for some river systems, even if not for the country as a whole. It would be useful if the WG report could be more explicit concerning this point.

The WG has previously discussed issues on nonstationary in the pseudo-stock and recruitment relationship. For most countries the S-R relationship is very weak resulting in the hockey-stick model selecting the min or max value previously recorded. If the time-series is reduced to the period since the decline in stocks (probably the 1990s), the time-series will become very short. There are also differing views about how changes in the S-R relationship should be accommodated in the CL estimation - i.e. if the decline in production is due to factors operating at sea, is it appropriate to change the CL for egg deposition, and hence production, in freshwater? Some countries that have developed river-specific CLs have already taken the change in marine mortality into account (e.g. UK (England and Wales)). However, we agree that this could be explored

Mortality associated with catch and release is not currently incorporated into the model. Data on $C \& R$ are currently incomplete. This could be explored further.

Most fisheries in home waters operate within rivers or estuaries and therefore catch salmon predominantly from a single stock. As a result, they can be targeted at stocks that are meeting their CLs - even if the national CL is not being met. Some countries permit fisheries on stocks that are not meeting their CLs (e.g. for socioeconomic reasons). In addition, many countries apply regulations on a multi-annual basis and take account of socio-economic factors in balancing the controls affecting different stakeholder groups and the rate of stock recovery that is planned. This means that fishing on some stocks may be permitted in years when these are not expected to meet their CL. This is an issue for managers and is being considered by NASCO within the current round of Implementation Plans.

| RG COMMENT | WGNAS RESPONSE |
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|  | This is discussed in Section 3.4 in relation to the <br> risk-based framework for provision of catch <br> advice for Faroes |

The RG understood that the run reconstruction model was run for each country (or region within country, e.g. in Russia or in UK(Scotland)) separately. In this way, a PFA was estimated for each country (or region within country), and the national PFAs were subsequently added to obtain a PFA estimate at the stock complex level. In the Monte Carlo procedure applied to estimate PFA for each country (or region within country), uncertainty in natural mortality was incorporated by simulating its value in each iteration from a Uniform $(0.02,0.04)$ distribution. If the intention of WG experts is that natural mortality in a given year should be the same for all countries, even if its value is uncertain, then the same value of natural mortality should be used in the same iteration for all countries. The RG understood that this is not done at present, treating each country separately in the run reconstruction model. This will probably lead to underestimating the uncertainty of PFA estimates at the stock complex level (because when adding up the PFAs of each iteration across countries, some countries will have a low value and other countries a high value of natural mortality in that iteration, hence cancelling the effects of a low or high natural mortality for all countries), potentially affecting conclusions about the stock complex status and ensuing management recommendations.

To determine if significant changes occurred in previously provided multiannual management advice, the WGNAS developed a Framework of Indicators (FWI). Within this FWI, a dataset was considered informative and kept when sample size was greater than or equal to 10 and $\mathrm{R}^{2}$ was greater than or equal to $20 \%$. It is unclear how a $\mathrm{R}^{2}$ of $20 \%$ would be considered informative? Prairie (1996. Evaluating the predictive power of regression models. Can. J. Fish. Aquat. Sci. 53: 490-492) developed a simple method to assess the predictive power of regression model based on the $R^{2}$ of the model. His analysis indicates that we could start distinguishing two groups when the $R^{2}$ exceeded approximately $60 \%$. Hence, the likelihood of arriving at a different conclusion using the current criteria of the FWI appears to be low, simply because a $\mathrm{R}^{2}$ of $20 \%$ is not very informative.
Consideration should be given to weighting the indicators taking their predictive ability into account, for example, based on the $\mathrm{R}^{2}$ of the relationship and averaging the weighted indicator states.

The same distribution of $M$ is used for all nation$\mathrm{al} /$ regional assessments. The ' R ' model originally used a different distribution of $M$ for each year (but not each country/region within each year), but this has now been brought in line with the CB model (i.e. use of a single distribution throughout).

The purpose of the FWI is to provide an initial appraisal of whether the previously provided multiannual management advice may have been incorrect. The Working Group recognizes that it would have been better to apply a more rigorous statistical level in selecting datasets for inclusion in the FWI. However, the decision to adopt a $20 \%$ was a pragmatic one, balancing strength of associations between indicators and PFA and a broad range of indicators over the four stock complexes. Use of a more stringent $\mathrm{R}^{2}$ criterion for selecting indicators to retain would have reduced the number of indicator stocks retained for the indicator (See Figure 3.7.2.1). As the PFA represents abundance integrated over a large number of river-specific stocks, the choice of criteria that resulted in a larger number of indicators per stock complex was favoured.

It was concluded to keep the criterion of $\mathrm{R}^{2} \geq 0.2$ in order to obtain a sufficient number of indicators to be able to use the FWIs even in the event of one or more indicators being unavailable at the time the FWIs are being applied. The $\mathrm{R}^{2}$ value of 0.2 corresponds to a value slightly lower than what is considered to be a "large" effect size ( $r=0.5, \mathrm{r}^{2}=$ 0.25 ) by Cohen (1988). Even though a criterion of $\mathrm{R}^{2} \geq 0.2$ gives each indicator little predictive power

| RG COMMENT | WGNAS RESPONSE |
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|  | alone (Prairie 1996), our approach with a suite of <br> indicators is more similar to meta-analysis <br> (Rosenthal 1984). Thus the outcome of the FWIs is <br> not dependent on the result of one indicator in <br> isolation, but rather on the combined perfor- <br> mance of the indicator set. |

Given that mixed-stocked fisheries can lead to the extirpation of weak stocks (Ricker. 1958. J. Fish. Res. Board Can. 15: 991-1006), alternative management options that minimize fishery impacts on weak stocks should be evaluated. A possibility could be assessing the spatio-temporal distribution of weak stocks in the Faroes Islands and West Greenland using DNA analyses. This may help to determine periods and areas that can be fished to avoid weak stocks in near real time (for instance, see Shaklee et al., 1999. Fish. Res. 43: 45-78).

Studies have been undertaken on the spatial and temporal distribution of salmon stocks from different countries at West Greenland and Faroes (e.g. Reddin et al., 2012 and Jacobsen et al., 2012). These studies indicate small differences in the distribution of different stock complexes (e.g. European fish tend to be caught further south than North American fish at West Greenland; and the proportion of maturing 1SW salmon from Southern Europe in catches at Faroes tended to be greater towards the south and early in the season). However, we do not know whether these trends apply to all river stocks. Given the very large number of stocks exploited in these mixedstock fisheries, substantial interannual variation in the areas prosecuted by these fisheries, and the wide distribution of weak stocks across the range of the species, the Working Group would not be able to provide advice to managers on differences in stocks being exploited at Faroes by season or other factors. DNA analysis of historic scale samples is currently underway to further investigate the composition of catches at West Greenland and Faroes, but this may not provide the level of resolution to identify stocks below the region level.

Management objectives should be clarified for mixed-stock fisheries, which take catch from different stocks. For a given probability " $p$ " (e.g. $p=0.95$ ), the objective (a) that "all stocks taken by the fishery are above conservation limits with probability $p^{\prime \prime}$ is less stringent than the objective (b) "that there is a probability $p$ that all stocks taken by the fishery are simultaneously above conservation limits". Depending on the degree of correlation between the status of different stocks caught in the mixed fishery, the two potential management objectives can be quite different. Thinking for the sake of argument in an equilibrium situation, objective (b) essentially means that there is at most a proportion 1-p of years where some stocks may fail to reach conservation limits, whereas objective (a) allows for some stocks to be below conservation limits in all years provided that the proportion of years below conservation limits does not exceed 1-p for any stock. Management objectives for individual stocks have often been expressed as $95 \%$ probabilities (i.e. using $p=0.95$ ). However, requiring $95 \%$ for objectives expressed simultaneously for a collection of stocks can be much more stringent. It is important that this issue is understood and clarified as appropriate. The RG noted that for the West Greenland Commission, NASCO has agreed to a

This is discussed in Section 3.4. WGNAS has agreed that the catch options table should show the probability that each management unit will achieve its CL/SER under different catch options. The WG considers that it is the responsibility of managers to determine the probabilities that they wish to adopt and they can then read off the appropriate catch option that meets this requirement. Managers may be more familiar with the concept of individual stocks (or stock complexes) achieving their reference levels than simultaneous attainment. However, the concept of simultaneous attainment may be informative to managers because 1-probability of simultaneous attainment quantifies the chance that one or more stocks or complexes in the coming year will fail to meet its conservation objective by chance alone, and not due to management failure or other changes in vital rates that may affect the abundance of salmon in a given year.

| RG COMMENT | WGNAS RESPONSE |
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$75 \%$ chance of simultaneously meeting or exceeding the management objectives. The use of the word "simultaneous", although not new in the way ICES has provided advice for the salmon fishery at West Greenland, is new in the context of stock assessments for other species. In a mixedstock fishery, even if advice is provided based on a probability $p$ (e.g. $p=0.95$ ) that each individual stock is above its conservation limit, it is nevertheless also informative to examine what this implies in terms of the probability that all stocks are simultaneously above conservation limits. The latter probability will decrease as the number of stocks exploited by the fishery increases. Hence, reducing the number of stocks exploited by the mixed fishery (e.g. by considering area and time specific management, if such information is available) is a way to increasing probability of the simultaneous event

## Section 4: North American Commission

| RG comment | WGNAS response |
| :--- | :--- |
| In the NEAC assessment all the data that were <br> used in the figures were readily available in The Working Group has addressed this issue. All |  |
| tables. Although this increased the length of the |  |
| chapter, it made it easier to compare the numbers in some cases as annexes. |  |
| presented in the text with those presented in the |  |
| figures. The RG recommends that a similar ap- |  |
| proach be adopted for the North American as- |  |
| sessment. In the latter, results were generally |  |
| presented only as figures or tables, but not as |  |
| both. |  |

Some of the information presented in Section 4.3.6 was already presented in previous sections (i.e. Section 4.3.2). Was this repetition necessary?

In the management advice (Section 4.4), WGNAS recommended habitat restoration in some areas as an alternative conservation action for wild populations that are critically low (Scotia-Fundy and USA). Although this is a common practice in several systems with stocks of conservation concerns, these measures have often failed to reestablish natural runs due to poor marine survival. Before recommending any costly restoration activities, it would be beneficial to estimate by how much freshwater survival/smolt output need to increase to achieve the target escapement goals. And then ask if these levels can be realistically attained?

The Working Group has addressed this issue.

The Working Group believes that freshwater restoration activities are commonly prioritized on a cost-benefit basis, but recognizes that more might be done to quantify the potential benefits that might arise from specific activities. The Working Group is aware that modelling approaches are being developed by a number of jurisdictions to assist this process. The dam impact analysis model for Atlantic Salmon in the Penobscot River, Maine (reported at Section 2.3 of this report) provides such an example.

The Bayesian inference and forecast model used to provide PFA forecasts (and used in the risk framework for the provision of catch advice) assumes heritability of age at maturity. Hence, the lagged eggs used in the model for the 2SW stock complex arose exclusively from 2 SW spawners.

The differences are due to convenience and determined by the age groups which are exploited in each of the fisheries. North American salmon are mostly caught at West Greenland and at the 1SW non-maturing stage (fish destined to become 2SW salmon). This contrasts with the Faroes

| RG COMmENT | WGNAS RESPONSE |
| :--- | :--- |
| The opposite approach appears to have been used <br> in the Northeast Atlantic, where lagged eggs | fishery where both 1SW maturing, 1SW non- <br> maturing, and 2SW fish are captured. For prag- |
| arose from both 1SW and MSW spawners. Clari- <br> fication on why different assumptions are made <br> for NEAC and NAC areas would be useful | me spawners was considered reasonable, and <br> the <br> consideration of two age groups was more ap- <br> propriate to the Faroes fishery. |

In the Bayesian inference and forecast model used to provide PFA forecasts, a time-varying productivity parameter to go from lagged eggs (or lagged spawners) to PFA is used. The actual modelling details are a bit different for NEAC (where each stock complex is modelled as a single unit) and NAC, where six management units are distinguished within the complex and this is taken into account in the specification of the productivity parameter. Nevertheless, the essential modelling mechanism is a random walk over time for productivity in both cases. In the forecasts, the uncertainty associated with productivity quickly increases as a consequence of the random walk assumption, which translates into large uncertainty in the PFA forecasts. Alternative modelling assumptions on productivity might lead to lower uncertainty in the forecasts, although it is not clear what alternatives might exist to the currently used random walk. Some alternative models may reduce uncertainty at the cost of increasing bias. Hence, if alternative models are developed, their performance relative to the currently used random walk should be explored. This could be done by applying each model in forecasts starting from some year back in the past and checking their relative performance based on the currently estimated PFA values for those past years.

A way to reduce the uncertainty in the productivity parameter would be by incorporating an auxiliary variable that can explain the annual variations in the productivity value. This is not obvious in the present structure that integrates the freshwater portion of the life cycle, where density-dependence is expressed, and the first year (post-smolt) survival at sea. The RG noted the proposed approach by ECOKNOWS to develop a full life cycle model that would separate these two phases of the life cycle and allow then an exploration of physical and biological covariates that could provide some correlation with variations in productivity. Although potentially useful in an inference sense, such covariates would only be useful for forecasting if the state of the covariates in the forecast years was known or could be equally forecasted. In that case the RG agrees with WGNAS that the models should appropriately incorporate the uncertainties associated with the covariates and the Bayesian framework developed for these models is the appropriate approach.

The issue of verifying by retrospective analysis the changes to model parameters and structures is noted.

We agree with the considerations noted by the Review Group on appropriately incorporating uncertainty in all the input variables to any model.

## Section 5: West Greenland Commission

## RG COMMENT

WGNAS RESPONSE

WGNAS documented that there were some issues in obtaining samples from Nuuk and argued that the Enhance Sampling Programme was compromised as $>20 \%$ of the fish harvested in West Greenland are landed in Nuuk. Although the RG agrees that obtaining unbiased samples from this area is desirable, the WGNAS should run a series of simulations to determine how critical this information is for their assessment. In particular, how precise do they need to be before it becomes a serious impediment to the interpretation of the data?

Based on the run reconstruction, it was concluded that none of the stated management objectives would be met in a mixed fishery off West Greenland, primarily because there was a low probability of achieving the conservation limits for most regions in North America (there were no issue for reaching the conservation limits of the southern NEAC stock complex). Are there alternative management scenarios that should be considered? For instance, is there a way to target this fishery on NEAC stocks, and thereby reducing the fishing pressure on North American stocks? This strategy was used successfully in the Chinook salmon fishery off western Canada (i.e. Beacham et al., 2008. North American Journal of Fisheries Management 28: 849-855).

For the MSW Southern NEAC stock complex, a discrepancy was detected between the probabilities presented in Table 3.4.1.1 under no fishing ( $87.2 \%, 88.5 \%, 88.5 \%$ ) and in Table 5.4.1 under no fishing ( $0.978,0.949,0.944$ ). Upon checking by the WG chair, it was found that the values in Table 5.4.1 were incorrect, due to having used a wrong SER value; a corrected table was provided for use in the ICES advice sheet. Please ensure that values are correct in future years and cross-check results between different tables and figures in the WG report. Additionally, it would help to clarify in Table 3.4.1.1 that e.g. the 2012/2013 Faroes fishing season corresponds to PFA in 2013 (2014 returns), whereas in Table 5.4.1 the 2012 catch year corresponds to the PFA in 2012 (2013 returns). A modified version of Table 3.4.1.1 was provided by the WG chair, clarifying the PFA year. This is useful for outside readers, so please make sure that this clarification remains in place in future years. Finally, the WG chair also explained that some minor numerical differences remained because the Faroes fishery risk analysis employed the PFA results from the WinBUGS run directly, whereas the West Greenland fishery risk analysis employed a (lognormal) approximation to the PFA distribution. It would be good if con-

WGNAS recognizes that sampling bias needs to be further explored, but since this was a nonassessment year this wasn't considered at the 2013 meeting. There are proposals to address this prior to 2014. The Working Group notes that there are significant quantities of reported harvest at Nuuk and division-specific trends in the biological characteristics. However, to minimize any potential bias, extensive sampling has occurred in areas to both the north and south of Nuuk, enabling a reasonable assessment of the biological characteristics of the harvest.

The Working Group has previously recommended changes in fishery management practices (mesh sizes) to modify the proportions of different stock complexes exploited by the fishery. The Working Group has also documented differences in the spatial and temporal distribution of fish from different stock complexes in the West Greenland area. However, there has been substantial variation over time and between years in the proportions of fish from different complexes / continents of origin taken in different areas. In addition, there is marked interannual variability in the distribution of catches and effort in the different NAFO areas in this relatively small subsistence fishery. The Working Group considers that these factors would make the development of effective alternative management scenarios particularly challenging

The Working Group has noted these errors and inconsistencies and will look to ensure that the clarification provided in 2012 is carried forward in future reports.

The WINBUGS issue may have arisen from modelling two PFAs in the Faroes fishery risk assessment and catch options whereas the Open BUGS predictions are for single PFAs. The issue is thought to have been addressed, as the same CODA output streams are used in each summary.
RG COMMENT WGNAS RESPONSE
sistency could be gained between the two analyses, presumably by using the results from the WinBUGS run in both cases.

## Follow-up comments from Review Group Chair

FOLLOW-UP COMMENTS FROM RG CHAIR
WGNAS RESPONSE

Section 3.6.3 of WGNAS 2012 report (Modelling approach for the Faroes catch options risk framework):

In the definition of Nw1SW, $\mathrm{pD}^{*}$ is defined as the "proportion of total catch that is discarded". Shouldn't this instead be: "proportion of catch that is discarded divided by proportion of catch that is landed"?

When harvest rates are calculated as $\mathrm{Hij}=\mathrm{Nwij} / \mathrm{S}$ (to calculate the probability that PFA is above the SER): Shouldn't the calculation take into account the M that occurs in the elapsed months between the Faroes fishery and homewaters fisheries? (at the moment it seems as though all catch is taken at the same time... ?)

The equation $[N w 1 S W=$ Nwtotal $x$ pA1SW + (Nwtotal $x \quad p D x 0.8)]$ should read $[N w 1 S W=$ Nwtotal $x p A 1 S W^{*}+\left(\right.$ Nwtotal $x p D^{*} /\left(1-p D^{*}\right) x$ 0.8)] and has been corrected in this report

M is taken into account but this was omitted from the model description. The whole catch option assessment is calculated back to the time of recruitment to the Faroes fishery (i.e. the dates to which the PFA and SERs refer). Thus the status of the stocks in each management unit (i.e. complex of country) is assessed by comparing the forecast PFA with the Spawner Escapement Reserve (i.e. the CL raised back to the time of the Faroes fishery). This is now included in the model description in Section 3.4.4

Catch options in Table 10.4.1 of 2012 West Greenland Commission advice sheet: As the catch in Greenland increases, the probability of the Southern NEAC MSW stock complex achieving the conservation objective hardly changes. What's the explanation for this? (has the 40:60 sharing arrangement between Greenland and NEAC areas, indicated in Sect 5.8.3 of WGNAS 2012 report, not been applied?)

The exploitation rate at West Greenland for the catch options current shown in Table 10.4.1 is very low. This is partly because a smaller proportion of the southern European stocks is actually available to the fishery than for North American stocks and also because the European stocks currently comprise only about $10 \%$ of the catch (these factors are, of course related).

NASCO agreed the $40: 60$ sharing arrangement for NAC stocks before the NEAC stocks were introduced to the assessment (in 200?). As the fishery cannot target NAC vs. NEAC stocks therefore a similar sharing arrangement formula was used in both cases. In the risk assessment, the sharing arrangement is taken on the total catch option for Greenland and then the raised catch is partitioned into NAC and NEAC fish based on the predicted proportion of the catch by continent of origin based on a uniform distribution of proportion NAC bounded by min and max values for the previous five years. That catch of NEAC fish is then subtracted from the PFA at that point in time and evaluated relative to the CLs for the fish discounted to the time of return to homewaters.

Since the Southern NEAC MSW stock complex is present both in West Greenland and Faroes, catch options in these mixed-stock fisheries should be calculated jointly rather than separately, correct? I

This is correct and has been considered by the WG, although as noted it is not a significant issue while ICES is advising that there is no catch option available for either fishery. Howev-

| FoLLOW-UP comments from RG chair | WGNAS RESPONSE |
| :--- | :--- |
| think the present calculation of West Greenland | er, formally incorporating this into the advice is |
| catch options assumes there is no catch of this | currently hampered by the lack of any feedback |
| stock complex in the Faroes and vice versa...? (of | from NASCO on the proposed establishment of |
| course, this is not a concern if the catch advice is 0 | the risk based framework for provision of catch |
| for both fisheries, but i imagine this matters if we | advice for Faroes. |
| start to give catch advice >0 at some point ?) |  |

Definition/calculation of CLs seems rather different in different places, see e.g. $2^{\text {nd }}$ paragraph in Section 5.2 of WGNAS 2012 report. My understanding is that the calculation based on the hock-ey-stick inflection point would be akin to what ICES calls Blim for other stocks (hence, a limit point to avoid with very high probability, at least $95 \%$ ), but in other cases the paragraph says that the CL corresponds, essentially, to Bmsy (?). If the latter is true, this would suggest that ICES would normally provide advice in line with being above CL with at least $50 \%$ probability (that's essentially what we do for other stocks, although the advice relates to Fmsy not to spawning targets, but we don't select a specific probability value above $50 \%$, unless managers indicate they require something else). The fact that the CL has been derived as akin to Blim in some cases, but akin to Bmsy in other cases, seems rather problematic when it comes to establish a required probability of being above it for giving catch advice. Of course, I understand the salmon situation is rather complicated, with all the different rivers, nations, management units, homewaters and mixed fisheries. Anyway, I'm just raising this for you to be aware of it and to see if WGNAS has any additional thoughts on this.

NASCO (1998) and ICES (1997) have agreed that the limit reference point for individual salmon river stocks (i.e. the CL) as defined is synonymous with Bmsy. Because this is a limit it should be avoided with a high probability. NASCO has also agreed that jurisdictions should use management targets (MTs) to help in maintaining stocks above the CLs; this is a target and so there should be a $50 \%$ probability of exceeding it.

WGNAS has previously (on more than one occasion) discussed this with the ICES Fisheries Sec and concluded that it was consistent with what ICES advises for some marine fisheries. In particular, the current ICES advice for shortlived stocks which uses a Bescapement reference point is relevant in the context of salmon as the majority of the spawners in most stocks in NAC and NEAC are from new recruitment. As such, the use of a Bescapement reference would be appropriate and there should be a low probability of falling below this reference point to preclude recruitment failure. Finally, NASCO fully understands and has endorsed the concept of conservation limits as currently defined. There will be some resistance and communication difficulties to changing the generalized approach for salmon because it is now embedded in the

## Annex 10: Technical minutes from the Salmon Review Group

- RGSalmon
- ICES HQ, Copenhagen, 22-25 April, 2013.
- Participants: Carmen Fernández (Chair), Carrie Holt (WGNAS reviewer), Kjell Leonardsson (reviewer), Tapani Pakarinen (WGBAST chair), Ian Russell (WGNAS chair), Henrik Sparholt (Secretariat), Jonathan White (WGNAS).
- Review of ICES Working Group on Baltic Salmon and Trout (WGBAST).


## General

The Review Group (RG) acknowledges the efforts expended by WGNAS in undertaking a substantial body of work and producing a thorough and informative report on the status and trends of salmon in the North Atlantic. The RG concludes that the information and analyses detailed in the WGNAS report provide an appropriate basis to respond to the NASCO request for advice.

The RG also acknowledges the detailed response to the previous year's RG technical minutes (provided as Annex 9 of WGNAS 2013 report) and appreciates its insightfulness.

As with the previous year, although for different reasons, the compiled report was achieved at a late date giving little time for review. This year this was largely due to the short time period between the WG and the RG/ADG. A fully compiled report draft for the reviewers in advance of the RG/ADG meeting (at least one week in advance) would facilitate the review and advice drafting process significantly. The RG considers that a minimum of two working weeks between WG and RG/ADG would be required in order to give the WG members and chair enough time to complete their report, while also ensuring that the reviewers have adequate time to prepare for the RG/ADG.

This review report merges the comments of all reviewers involved in the process this year. The review focuses on three sections of the WGNAS report, pertaining to the Northeast Atlantic Commission area, North American Commission area, and the West Greenland Commission area (with an emphasis on the first).

The remaining sections of the WGNAS report provide a general introduction, an overview of salmon catches, landings and tag releases, and descriptions of new or emerging threats, successes and failures of restoration projects, threats of exotic species, Expert Group reports relevant to North Atlantic salmon, data deficiencies, monitoring needs, and research requirements. This was presented by the WG chair during the RG meeting, so reviewers are informed and aware, but no specific review of those aspects was undertaken by the RG.

## Comments by the reviewers on the WGNAS Report

## Section 3: Atlantic salmon in the North-East Atlantic Commission Area

1 ) In general, the analyses are technically correct, and the scope and depth of those analyses are appropriate to generate advice required. The comments pertain to aspects of the analyses that likely would not result in a change in advice for the current period, but may be useful in future.

2 ) Section 3.1.6 describes a general downward trend in proportion of 1SW salmon in the reported catch, especially in the Northern NEAC areas (since ~2005), and with country-specific variation. The text notes that the causes are uncertain, but may be due to management measures (e.g. resulting from size-selective fishing?). A similar trend is shown in the reconstructed spawner numbers for Northern NEAC, for which numbers have increased since 2005 for MSW spawners and remained stable for 1SW spawners. Could this trend be a result (at least in part) due to increases in proportion of fish maturing after MSW (due to, for example, a change in marine environmental conditions)? To what extent might continued trends in the proportion of fish maturing at each age (PFAm, pg. 116, Section 3.5.1) affect forecasts of PFA in the Risk Framework (which currently assumes constant mean PFAm 2012-2016 at 2011 levels)? See also comment 9 below.

WG response at the RG-meeting: WGNAS is aware of the changes in age composition of stocks and agrees that this could reflect a variety of factors including changing environmental conditions at sea and management actions. Need to check this in future.

3 ) Country-specific CLs depicted in Figures 3.3.4.1 (a-j) are based on residual sums of squares estimate of the hockey-stick model. For many countries (regions within countries) the CLs are near (at) the low end of estimated lagged egg abundances, suggesting the CLs may be overestimated. Although this is precautionary from a conservation perspective, it may result in unnecessary fishery closures if those countries/stocks constrain a multistock fishery. Indeed, the uncertainty in dropping below CLs is due to both uncertainty in current egg (or spawner) numbers and uncertainty in the CL itself. Have uncertainty estimates for CLs been considered (e.g. derived from the likelihood profile for the CL)? In addition, the acceptable buffer between current egg (or spawner) numbers and CL may depend on our certainty in the CL itself. For highly certain CLs, the buffer described on p. 99 (Section 3.2) that is derived from confidence intervals of spawner estimates may be sufficient. For highly uncertain CLs, a larger buffer may be prudent.

WG response at the RG-meeting: CL being used as fixed, but uncertainty could be included in future. Possibly a management issue, whether they want a fixed CL or uncertain CL's. Suggested to be brought up in the WG for discussion on how to deal with this.

* Comment from RG chair: I do not immediately follow the comment that CLs derived in this way may be overestimated. It would be good if WGNAS could clarify.
* Response from reviewer: My comment simply pertains to the observation that CLs are defined at the lower boundary of lagged egg abundances, where there may not be any evidence for reductions in PFA (e.g. Figure 3.3.4.1.f). But, I note the comment at the end of this section that CLs may be underestimated when derived for multiple asynchronous stocks within a region.
4 ) The assessment of spawner number against CLs and PFA against SERs give inconsistent results in some years (Figures 3.3.4.1 (a-j)). Can these differences be explained? Is there a reason why both are presented if CLs are typically the basis for management advice? I assume this is because the Bayesian forecasting model within the Risk Framework provides PFAs which are evaluated against SERs.

WG response at the RG-meeting: Both PFA/SER and Spawners/CLs are needed; the former is aimed at assessment of stock status at sea to inform management decisions on the
high seas fisheries; the latter provide an assessment of national stock status for individual countries/regions. The differences noted between the two reflect exploitation in homewater fisheries. The WG recognize that there is now some duplication of abundance in the summaries of stocks (Figure 3.3.4.1a-j) and the new country forecasts (Figures 3.5.4.13.5.4.11). Duplication may be skipped in future.

* Comment from RG chair: (just a minor addition to this discussion, in case it helps to clarify the point about "inconsistency") My understanding is that SERs are the CLs (which refer to spawning time) corrected for the natural mortality that occurs between PFA time (before fishing exploitation begins) and spawning time. Hence, it may well happen that the PFA is above the SER (as this is before exploitation begins) but the number of spawners is below the CL (as this is after exploitation has occurred). If the PFA of a stock complex is found to be below the SER in a given year, no mixed-stock fisheries are possible on the complex.
5 ) The CLs for Ireland changed significantly from 2012 to 2013 due to a change in methodology. Presumably the revised set of stocks used in the Bayesian hierarchical analyses conform to the exchangeability assumption to a higher degree than the prior set, but it is not possible to evaluate this without relevant model and data (but perhaps this is outside the scope of the current report?).

WG response at the RG-meeting: River-specific CLs were re-calculated in 2012-2013 for Irish national stock assessments. The recalculated CLs were based on up to date biology and catch data. The stock-recruitment approach used was not the hockey-stick approach used for other regions of the NEAC assessment, which may impose some inconsistencies with the country/stock-complex analyses based on the hockey-stick approach.

* Comment from RG chair: As a general point (not specifically for Ireland), it would be good if WGNAS could provide some more background information on how CLs are computed. During RG discussions it was not possible to clarify completely how CLs had been derived, when based on an MSY concept. Clarification of this, and inclusion in the Stock Annex, would help.
* It was discussed during the RG meeting whether the hockey-stick approach used sometimes at the national level is consistent (or inconsistent) with the approach of having riv-er-specific CLs (then summed up to national level) based on, we understand, different stock-recruit relationships (e.g. Ricker). Does this matter for the consistency of the results? (Please see also additional comment on stock-recruit analysis discussed at the end of this report). In essence, there are two questions here: (1) one refers to computing CLs by river and then summing up to national level vs. computing CLs directly at national level; (2) the second question is about the potential impact of using alternative stockrecruit forms (e.g. hockey-stick vs. Ricker or Beverton-Holt).
6 ) Although the updates to the run-reconstruction model seem reasonable (Section 3.3.3), the model itself is not provided, so cannot be reviewed (and has already been reviewed). It is noted that "errors in the outputs largely reflect uncertainties in the estimates of the data". One way to account for uncertainties in model input is a state-space model that explicitly considers errors in the data (for a Pacific salmon example, see Fleishman et al., 2013). Without reviewing previous WGNAS reports for a model description, it's difficult to assess to what extent that approach would be useful (or is already implemented).
6.1) Interestingly, the size of the confidence limits on spawner numbers will depend on the extent to which uncertainties are considered in
the model, which has implications on stock status relative to CLs. The more assumptions made in the model, the smaller the size of the confidence intervals and the smaller the buffer (and vice versa).

WG response at the RG-meeting: The model is mainly an accounting tool. Uncertainties (max-min expert estimates of the exploitation). That is not so much model errors as uncertainties from input estimates.
Ambition to provide a Stock Annex in future versions of the report, where methodology will be detailed or pointers given for finding the appropriate method description.
6.2) When plotting results for exploitation rates of 1SW and MSW (Figures 3.3.4.1 a-j), I suggest 'jittering' the data points so that both 1SW and MSW points and confidence intervals can be viewed in each year.

WG response at the RG-meeting: $O K$, will be done.
7 ) The report correctly notes that management objectives are required to proceed in providing useful advice for management, and this point cannot be overemphasized. Indeed, for the Risk Framework, management objectives would inform the choice of management unit and share arrangements (Section 3.4.1., p.107). It is noted that NASCO's recommendation to base fisheries decisions on river and age-specific CL's is contradictory to NASCO's agreement to manage distant water fisheries on four stock complexes in NEAC (which are much larger than those in the West Greenland fishery). Provisionally (?), the choice to provide management advice at the stock complex level, and provide implications of that advice at the country level seems like a pragmatic approach. Indeed when applied to the Risk Framework, these approaches give consistent catch advice (fishery closure), but this may not be the case in future. Given the possibility of future assessments at river-specific level, it may be necessary to derive more sophisticated management approaches that incorporate emerging information on stock identification and stock-specific spatial and temporal migration patterns (e.g. to avoid exploitation of weak stocks through spatial/temporal fishing restrictions).

WG response at the RG-meeting [N.B. the reviewer indicated this was intended as more of a comment than a suggested action]: NASCO is aware of the difficulties of extending the river-specific approach to the management of mixed-stock high seas fisheries where over 1000 individual stocks can be exploited, and have accepted that management decisions on these fisheries should be based on assessment of stock complexes. The WG noted the comments. The WG also advised that it has previously advised on the use of different management strategies to target specific stock complexes in high seas fisheries (e.g. changes in mesh sizes) and has documented spatial and temporal differences in the distribution of fish in the sea, which might provide a basis for future management. The WG intends to continue to incorporate emerging information on stock identification/distribution in its advice.

8 ) Choice of risk levels is recommended on p. 112 (Section 3.4.3), but would this depend on trade-offs between values derived from the fishery as an aggregate and value of conserving a diversity of stocks? In Canadian Pacific salmon fisheries, such decisions typically require engagement of stakeholders to include societal values.

WG response at the RG-meeting [N.B. the reviewer indicated this was intended as more of a comment than a suggested action]: The WG noted that NASCO recognizes the need for socio-economic factors to be considered in management, but that this was not considered explicitly in their work. It was noted that societal values were often acknowledged in management (e.g. Aboriginal and food fisheries in Canada).

9 ) The Risk Framework includes a Bayesian forecast model that generates forecast for PFA generated in WinBUGS (Section 3.4.4). It's not clear from the text which parameters were given priors, which prior distributions were used, and the impact of those priors on the posteriors (though perhaps this is described and reviewed in a previous WGNAS report?). In addition, the structure of the model is unclear. Are the 1SW (maturing) and MSW catch covariates included in the model to predict PFAt, as in the equation on p.115? 1SW non-maturing fish not considered in the model (p.116)? Is this because they are not caught in Faroes fishery?
9.1) The productivity parameter is derived independently for each stock complex and/or country. However, given similar trends in marine survival among stocks from countries, there may be value in developing a hierarchical model that estimates productivity parameters from a shared hyper-distribution among those groups. Has this been considered?

Complexes were separated as the development in the productivity diverged for the different complexes. Hyperdistribution should be considered.
9.2 ) In addition, the forecast component of the model assumes constant average productivity over time (2012-2016), despite evidence of declining marine survival. Additional sensitivity analyses could show probabilities of achieving CLs (and associated catch implications) from different assumptions about a continued decline in productivity vs. constant productivity (as similar approach has been applied to Pacific salmon on the Fraser River, Canada), as well as a continued trend in PFAm and constant average PFAm (see also comment 2 above).

WG response at the RG-meeting: Valuable comments. WG provided some explanations about the forecast model, and agreed it might be useful to reexamine the productivity parameter and to explore additional sensitivity analysis.

* Comment from RG chair: I agree with the reviewer's general comment (before getting into parts $a$ or b) that the description of the Bayesian forecast model needs some improving, in terms of what data are being used as "data", which parameters are given prior distributions, what the observation equations are (i.e. how are the observations linked to the underlying model variables and parameters), etc. This will hopefully be addressed as part of the Stock Annex.
* Related comment from RG chair: it seems to me that the Bayesian forecast model could actually be made into a closed loop, so that the whole cycle from lagged eggs to PFA, returns, spawners, and again lagged eggs, could be modelled consistently in a loop (without running a separate run-reconstruction model). It would be interesting to get WGNAS views on this.
* Response from reviewer: yes, I think this approach would provide an opportunity to account for uncertainties in the run-reconstruction model in a more realistic way.
10 ) The Risk Framework assumes monthly instantaneous mortality of 0.03 (Section 3.4.4., p.114). How was this derived? What are the implications of assuming (more realistic) variability in this value?
10.1 ) A major assumption in the Risk Framework applied at the country level is the apportioning of catches to management units (Section 3.4.5). The text states that an alternative method was proposed for estimating the split of catches in 2012 (p. 115), but the results of that alternative method are not described here. Given continuing lack of fishery derived data around the Faroes (due to lack of a fishery), pelagic fishery bycatch of salmon could become an important source of stock- or country-specific information, if those fish can be identified to country/stock, as noted in the text. This opportunity should be emphasized.

A similar question was asked by the RG in 2012, and the response from the WG can be found in Annex 9 of this year's report. Mortality at sea is an ongoing issue and efforts are being made to better partition the at sea mortality. The issue has been subject to detailed investigation in the past and 0.03 identified as the most suitable current value. Its continued use was essentially a pragmatic decision; in the absence of suitable information to vary this parameter there was no basis to change it. The alternative method for apportioning catch referred to was a genetic analysis of historic scales. This work had not proceeded as well as possible (degradation of DNA in the samples) but work was continuing and it was hoped results would better inform future management.

A few additional comments on the risk framework for catch options at the Faroes:

- The exploitation rate for maturing 1SW (from both Northern and Southern NEAC) salmon at the Faroes seems very low, and this raises the question of whether these two stock complexes should be included in the risk framework for the Faroes. At the moment, their inclusion does not affect the catch advice for the Faroes (which would be zero in any case, given that the PFA of the Southern NEAC MSW stock complex is below the SER). But their inclusion, if not needed, could lead to unnecessarily restrictive advice for the Faroes fishery in future. The RG requests WGNAS to consider this question in their next meeting.
- In Table 3.6.1.1 it should be made clear the years in which potential returns are being measured against CLs. Because the Faroese fishery seems to exploit mainly MSW salmon during their second winter in the sea (so fish that are due to return just after the Faroese fishery takes place), the RG understands that it would make most sense to measure, e.g. for catch options in 2013/2014, the potential returns in 2014 vs. the CL in 2014. This should be made clear in the presentation of Table 3.6.1.1.

11 ) Section 3.7.2 recommends that the framework for indicators approach be revised so that an assessment for a closed fishery is only triggered when the indicators are above the upper $75 \%$ confidence limit. This is a reasonable recommendation given finite resources for assessments.

WG response at the RG-meeting: Welcomed supportive comments.
12 ) The R software provides a less error-prone platform (though not errorfree!) for performing statistical analyses that involve multiple datasets than Excel spreadsheets that usually require multiple cutting and pasting steps. There may be value in transferring the analyses for the framework of indicators (Section 3.7.2) from spreadsheets to R , and providing R code in the annex (for this analysis, and other models) to this report for review.

WG response at the RG-meeting: Parallel runs (Crystal ball and R) for control purposes this year, but the ambition is to shift to $R$.

## Minor comments on formatting

1 ) The inclusion of Equation numbers would aid in the review process when comments refer to specific parameters or equations.
WG noted the suggestion. Future development of stock annex should help with this.
2 ) Annexes that include model description, equations, and R/WinBUGS code for all models would also help in the review process. Such annexes could be appended annually to each assessment (or for years when an assessment is performed). Although a folder for "software" was noted on the WGNAS website, it did not contain any code (as of Wednesday 17 April).

WG noted the suggestions. The code was available on the WGNAS SharePoint site but in a different location. Future development of stock annex should help with this.

3 ) At least two tables were misplaced and Figures were commonly cut off of the printed page. This is a common consequence of managing such a large file in MSWord, when figures and tables are pasted from different software packages (e.g. Excel). An alternative software for developing complex documents such as this one, is LaTeX, which can be seamlessly be integrated with R code to create figures, tables, and captions that are incorporated with the text with user-specified formatting. LaTeX is commonly used for the documentation of complex stock assessments in Canada, and is favoured, in part because assessments can be updated with additional data in subsequent years very easily ("with the click of a button") since figures and tables are automatically generated.

Something for ICES to consider.

## Section 4. North American Commission

1 ) The NASCO Framework of Indicators for NAC indicated that an evaluation of catch options and management advice were not required. The assessment was updated with 2012 data, but the modelling approach remains unchanged from previous years, and therefore is not reviewed here.

2 ) The assessment of continued low abundances of stocks across North America (especially in USA and Scotia-Fundy areas) is supported by the updated data. As noted in the text, given the consistent declines over
broad spatial scales, reductions in marine survival for selected stocks where monitoring exists, and sustained smolt production over time, it is likely that this depletion is due in large part to factors acting on marine survival in the first and second years at sea.
3 ) However, several gaps in data are noted. First a change in monitoring of adult returns in Labrador from four counting facilities to only three in 2010 and 2012 may have caused the large variability in returns (especially for large returns) in the last several years (Figure 4.3.2.1 and 4.3.2.2). The previous time-series could be re-analysed omitting information from the 4th counting facility to identify if variability in the last few years are from change in monitoring, or are driven by population dynamics. I suggest highlighting those years in the Figures to emphasize the possible different interpretation of those values. Given this issue, and the large area covered by a single counting facility in SFA1, I agree with the authors that, "Future work is needed to understand the best use of these data in describing stock status and the Working Group recommends that additional data be considered in Labrador to better estimate salmon returns in that region" (p. 225).

WG response at the RG-meeting: The loss of one monitoring facility was a temporary problem (loss of trapping facility due to flooding). The absence of this facility was unfortunate, but it was not considered to explain the large variability of the returns in Labrador which were consistent with other parts of North America and indicative of wider coherent issues acting in the sea on stocks from a broad geographic area. Thus, other big changes in the region as a whole (North America) indicates that the variability could be explained without addressing the uncertainty/variability.

4 ) The section on the estimates of total abundances for Scotia-Fundy states that the current model overestimates total abundances. It's unclear whether this overestimate is only for the current year (2012), or for the entire time-series. Given the dramatic declines in 2012, I have assumed they are for the entire time-series. In addition, I suggest including the ranking of abundances in this section to emphasize that for several time-series the current abundances are the lowest on record.

WG response at the RG-meeting: Will add a line or two in the report to clarify this issue. Clarification has already been sought from Canadian colleagues and clarification will be included in 2013 report before it is finalised.

5 ) In Section 4.3.4. (Egg deposition), for what portion of rivers have CLs been identified?

WG response at the RG-meeting: CLs are only presented for 74 (of $\sim 1000$ rivers) where detailed monitoring takes place, although CLS have been determined for over 400 (ca. $40 \%$ ) Canadian rivers many of which are relatively small (CLs of around 200-300 spawners).

6 ) In Section 4.3.5. (Marine survival, return rates), the declines in marine survival in 2012 from 2011 are alarming at first, but are in large part due to relatively high marine survival in 2011. Five-year average analyses provide more meaningful results.
6.1) Are the declines in Gulf region significant? Results are not provided in the text, but are presented in Figure 4.3.5.1.

WG response at the RG-meeting: Will add a line or two in the report to clarify this issue. Clarification has already been sought from Canadian colleagues and clarification will be included in 2013 report before it is finalised.

## Section 5. West Greenland Commission

1) The NASCO Framework of Indicators for NAC indicated that an evaluation of catch options and management advice were not required. The updated 2012 assessment is based on status of stocks in the NEAC and NAC (reviewed above).
2 ) Additional information on the number of NAC and NEAC salmon caught in West Greenland (Figure 5.1.3.2.) is provided to estimate impact of the West Greenland fishery on those stocks. Currently, sampling to determine continent of origin is based on three sampling stations (omitting sampling station at Nuuk), that do not cover the spatial range of the fishery. The report notes that the lack sampling at Nuuk compromises the ability to correctly identify biological characteristics of the catch (including continent of origin). However, the figure depicting temporal trends in catch of NEAC and NAC salmon (Figure 5.1.3.2) does not include confidence limits, so the consequences of increased uncertainty in biological characteristics are not shown. If included, a large increase in the range covered by the confidence limits in 2012 due to a reduction in information about continent of origin might clarify the importance of those samples.

Response by the WG: This question is similar to that asked in the last year's TM, see Annex 9 in the report.

WG response at the RG-meeting: Samples are also collected in the surrounding region so there are data to provide a reasonable description of stock composition both temporally and spatially. However, the WG continues to note this issue and to recommend that action is taken to resolve the difficulties at Nuuk. The WG will consider the possibility of including confidence intervals in the figure in future.

General comment on the use of a single stock-recruit function to represent an entire stock complex or stocks from many rivers at the nation level.

The stock-recruit functions used for salmonids generally includes densitydependence in the freshwater environment. Since the density-dependence is a local process caused by for example competitive interactions it is difficult to justify a region wide outcome, especially since the different river stock sizes may not cover completely. It can easily be verified that the sum of several local stock-recruit (SR) functions cannot be reformulated in to a single function with the same few parameters. For example, joining two Hockey-stick SR-functions would imply (focusing only on the linear parts):

$$
\mathrm{R} 1+\mathrm{R} 2=\mathrm{a} 1^{*} \mathrm{E} 1+\mathrm{a} 2^{*} \mathrm{E} 2
$$

where R denotes recruits and E denotes eggs.
Merging these into a single function:

$$
\mathrm{ax}^{*}(\mathrm{E} 1+\mathrm{E} 2)=\mathrm{a} 1^{*} \mathrm{E} 1+\mathrm{a} 2^{*} \mathrm{E} 2,
$$

requires that ax will be a function of a1, a2, E1 and E2;

$$
\mathrm{ax}=\left(\mathrm{a} 1^{*} \mathrm{E} 1+\mathrm{a} 2^{*} \mathrm{E} 2\right) /(\mathrm{E} 1+\mathrm{E} 2),
$$

that is, ax is not a constant! In contrast the maximum threshold will sum up for all rivers, but the question is how often that limit will be reached when joining data from many rivers? In the WG analyses the statistical fit of the upper limit to the data, but the question is how relevant that is.

Statistically it will still be possible to fit a traditional SR-curve to multistock data, but there will most likely be an additional level of uncertainty (due to the dependence of the parameters ax on how the number of eggs are distributed among the various populations). Fitting the entire hockey-stick, that is, also to the maximum recruitment, to the data are likely to lead to underestimation of the maximum recruitment capacity. Independent measures of the maximum capacity as that based on the number of recruits related to the wet area of the rivers should be considered for all regions. When the maximum recruitment capacity depends on a statistical fit one needs to assume that all or most stock dynamics are synchronous. If this is not the case the maximum recruitment is not likely to be covered by the estimated or observed spawner counts. Consequently the maximum recruitments are likely to lie above the currently fitted maximum recruitment lines. If this is the case, then there is a risk that the current CLs are underestimated. The RG has no solution to suggest on how to solve this issue (besides using the wet area approach).

## Reference Cited

Fleischman, S.J., Catalano, M.J., Clark, R.A., and Bernard, D.R. 2013. An age-structured statespace stock-recruit model for Pacific salmon (Oncorhynchus spp.) Can. J. Fish. Aquat. Sci. 70. dx.doi.org/10.1139/cjfas-2012-0112.


[^0]:    ${ }^{1}$ Lower Adour only since 1994 （Southwestern France），due to fo fishery closure in the Loire Basin．
    Adour estuary only（Southwestern France）．
    Number of fishermen or boats using drift nets：overestimates the actual number of fishermen targeting salmon by a factor 2 or 3 ．
    Common licence for salmon and sea trout introduced in 1986 ，leading to a short－term increase in the number of licences issued．
    Before 2000，equal to the number of salmon licenses sold．From 2000 onwards，number estimated because of a single sea trout and salmon angling license．
    （2012／mean－1）${ }^{*} 10$

[^1]:    MSW(min)

