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## Executive summary

## Faroe Bank Cod

The fishing area has been closed since 2009, and total reported landings in 2015 were the lowest recorded since 1965 ( 17 tonnes).

Spring survey index suggests that the stock increased from 2012 to 2014 and declined substantially again in 2015 and 2016. Both summer and spring index suggest that the stock size has been well below average since 2004, and there are no indications of strong incoming year classes. Since 2008 the stock is mostly comprised of large individuals ( $>80 \mathrm{~cm}$ ). Correlation of recruitment year classes between the surveys since 1995 is $\mathrm{R}=0.86$.

The advice is that no fishing effort should take place on this stock until significant rebuilding has taken place.

## Faroe Plateau cod

The input data in this update assessment consisted of the catch-at-age starting in 1959 and spring survey starting in 1994 as well as summer survey starting in 1996. The maturities were obtained from spring survey. The terminal year in the assessment was 2015.

The assessment settings were the same as last year. An Extended Survivor Analysis was tuned with the two survey indices. The fishing mortality in the terminal year was estimated at 0.46 , which was higher than the Fmsy of 0.32 . The total stock size at the beginning of the terminal year was estimated at 28500 tonnes and the spawning-stock biomass at 19700 tonnes, which was slightly below the limit biomass of 21000 tonnes. Based on preliminary analyses, the extremely low biomass since 2005 seems to be unprecedented in the last three centuries.
The short-term prediction until year 2018 showed a slightly decreasing total-stock biomass to 25000 tonnes and a spawning-stock biomass to 19000 tonnes. It is advised to reduce the fishing mortality to the lowest possible level to rebuild the stock. This stock is scheduled for a benchmark in 2017.

## Faroe haddock

Being an update assessment, the changes compared to last year are additions of new data from 2015 and 2016 and some minor revisions of recent landings data with corresponding revisions of the catch-at-age data. The main assessment tool is XSA tuned with 2 research vessel bottom-trawl surveys. The results are in line with those from 2015, showing a very low SSB mainly due to poor recruitment but also due to higher than recommended fishing mortalities in recent years. SSB is now estimated below Blim, but with an improved recruitment, especially of the 2015 YC, combined with substantially earlier maturity-at-age, it is predicted to increase to just above Blim in 2017, and be above $B_{p a}$ in 2018 - with the assumption of status quo fishing mortality in all years. Fishing mortality in 2015 is estimated at 0.26 and the average fishing mortality 20132015 at 0.27 (FMSY and $\mathrm{F}_{\mathrm{pa}}=0.25$ ). Landings in 2015 were only 3395 t , slightly higher than in 2012 to 2014. This year's assessment indicates that the 2015 assessment underestimated the 2014 recruitment by $42 \%$ ( 2.6 million vs. 4.5 million, , overestimated the fishing mortality in 2014 by $10 \%$ ( 0.26 vs. 0.29 and underestimated the 2014 total- and spawning-stock biomasses by $7 \%$ and $27 \%$, respectively ( 20 and 18 thous. t vs. 18 and 14 thous. t). This stock is scheduled for a benchmark in 2017.

## Faroe Saithe

Nominal landings in 2015 are estimated at 25 kt. Estimated fishing mortality in 2015 (average of ages 4 to 8 ) is $\mathrm{F}=0.25$, which is lower than the historical average ( $\mathrm{F}=0.35$ ) and below $\mathrm{F}_{\mathrm{MSY}}=0.30\left(=\mathrm{F}_{\mathrm{pa}}\right)$. Due to high fishing mortality SSB decreased substantially from 2005 to 2013 but it increased again in 2014 and 2015 as a consequence of low fishing mortalities, improved weights and increasing maturity ogives.
Numbers of the most recent yearclass (2012, age 3 in 2015) is estimated at 63 mill. A statistical separable model (used as a diagnostic tool to the spaly assessment) suggests that recruitment in 2015 is not as strong as the spaly assessment estimate and it predicts recruitment for 2015 at 36 mill. Retrospective plots indicate an overestimation of yearclass strength in recent years.

With a status-quo $F_{\text {bar }}(2016)=0.25$ and recruitment of 30 mil ., the SSB is predicted to increase to 97 kt . in 2016.

This stock is scheduled for a benchmark in 2017.

## Icelandic Saithe

The 2016 reference biomass ( $B_{4+}$ ) is estimated at 273 kt , around the average in the assessment period (1980 to the present). Spawning biomass is estimated as 139 kt , above the average in the assessment period and well above $B_{\text {trigger }}=65 \mathrm{kt}$ and $B \lim =44 \mathrm{kt}$.

Harvest rate has been below the target of 0.2 in last two years and fishing mortality is also predicted to be low. The reason is that the TAC has not been caught, most likely a problem of availability but a smaller than predicted stock cannot be excluded.

Weights of ages 3-6 have been low in recent years, but older ages are close to average weight. Maturity-at-ages 4-9 has decreased in recent years and is currently around average.
Recruitment has been above average since 2009 and relatively stable. There are indications from the survey in 2016 that the 2012 yearclass might be strong.

The assessment model is a separable statistical catch-at-age model implemented in AD Model Builder. Selectivity is age-specific and varies between three periods: 1980-1996, 1997-2003, and 2004 onwards. The result of the assessment is relatively insensitive to settings of the assessment model and resulting reference biomass in 2016 varies from 252-310 thous tonnes with the lowest values obtained from models with random walk constraint on fishing mortality. The assessment is considered relatively uncertain but this year's assessment introduces slight downward revision from last year's assessment.

In 2013, the Icelandic government adopted a harvest control rule for managing the Icelandic saithe fishery; evaluated by ICES (2013). It is similar to the $20 \%$ rule used for the Icelandic cod fishery. When the population is above $B_{\text {trigger, }}$ the TAC equals the average of $0.2 B_{4+}$ in the assessment year and last year's TAC.

According to the adopted harvest control rule, the TAC will be 55 kt in 2016/2017, the same as in the current fishing year.

## Icelandic cod

A formal HCR has been in place to determine the TAC for this stock since 1994. The rule has gone through amendments and revisions. The last significant change occurred
in 2007, when the harvest rate multiplier upon which the TAC for the next fishing season is based was changed from 0.25 to 0.20 of the reference biomass $\left(B_{4+}\right.$ - biomass of age 4 years and older) as estimated at the beginning of the assessment year. According to this decision rule the TAC will be 244 kt in the 2016/2017 fishing season. ICES evaluated the plan in 2009 and concluded that it is in accordance with the both the precautionary approach and the ICES MSY framework.

The decision on the TAC each year is based on an analytical stock assessment carried out each year. The results from the assessment show that the spawning stock in 2016 is estimated at 469 kt . The values estimated in recent years are higher than has been observed over the last five decades. The reference biomass $\left(B_{4+, 2016}\right)$ is estimated to be 1243 kt , and has not been so high since the late 1970's. Fishing mortality, being 0.27 in 2015, has declined significantly in recent years and is currently the lowest observed in the last six decades. Year classes since the mid-1980s are estimated to be relatively stable but with the mean around the lower values observed in the period 1955 to 1985. The first indication of year classes 2014 and 2015 indicate that they may be larger than that observed in the recent decades.

Mean weight at age in the stock and the catches that were record low in 2006-8 have been increasing in recent years and are now around the long term mean. Spring survey weights in some of the important age groups have, however, declined in 2016 compared with 2015. Catch weights are hence estimated to be lower in 2016 compared to 2015.

The inputs in the analytical assessments are catch-at-age 1955-2015 and spring groundfish survey (SMB) indices at age from 1985-2016 and fall survey groundfish survey (SMH) indices at age from 1996-2015. The Catch-at-age Model (ADCAM) has been used as the basis for the advice since 2002 (spring survey only up to the 2009 assessment, both surveys since then).

## Icelandic haddock

The 2014 year class is estimated to be large, after six consecutive small year classes from 2008-2013. The 2015 yearclass is expected to be close to the geometric mean. The current assessment shows some upward revision of the stock compared to last year's assessment. The main features are though the same; that the fisheries are currently mostly based on relatively small yearclasses. It is not until 2018 that the 2014 yearclass affects the fisheries.

Growth in 2015 was above the 1985-2015 average but less than predicted. The mean weight of young fish is above average while old fish are close to average. The assessment procedure was the same as last year (SPALY), an Adapt type model tuned with both surveys.

There are differences in the perception of the state of stock in assessments based on either spring or autumn survey, with autumn survey indicating a larger stock. It has been like that since 2009. Different models using the same tuning data show similar results.

Advice is given according to the adopted Harvest Control Rule, and the advice for the fishing year 2016/2017 is 34600 tonnes.

No environmental drivers or ecosystem effects are known that can help in prediction of the development of the haddock stock. Some effect of the environment on the stock can though not be excluded.

## Icelandic summer spawning herring

The total reported landings in 2015/16 fishing season were 69.7 kt (including summer fishery 2015) but the TAC was set at 71 kt -the difference being transfer of quota between years. The fishable stock (age $4+$ ) in winter surveys 2015/16 was estimated at 372 kt , compare to 433 kt in winter 2014/15. The 2014 year class (age 1 in 2015) appears small.

This is an update assessment where the 2015 data have been added to the input data and no revisions of last year's data. The analytical assessment model, NFT-Adapt, indicates that the biomass of age $3+$ will be 393 kt and SSB is 318 kt at the spawning time in 2016. Continuation of relatively small year classes entering the spawning stock causes a further decline in SSB even if still above $\mathrm{B}_{\text {pa }}$ of 300 kt . Fishing at $\mathrm{F}_{\mathrm{ms}}=\mathrm{F}_{0.1}=0.22$ in the fishing season 2016/17 will give a catch of 63 thousand tons resulting in SSB in 2017 expected to be 303 kt .

There are indications for new infection by Ichthyophonus in the stock for the second year in row even if all data support that additional natural mortality in the stock due to the infection should only be applied for the first two years of the main outburst that started in 2009. This calls however, for continuation of the extensive monitoring and studies.

## Capelin in the Iceland-East Greenland-Jan Mayen area

In May 2015 ICES advised that the initial (preliminary) quota should be 54000 t . In June 2015 the coastal states agreed on an initial quota of 300000 t . In October, the Icelandic Marine Research Institute (MRI) declared an intermediate TAC of 44000 based on an acoustic survey in September and the HCR from ICES WKICE (2015). Lastly, a final TAC of 173300 t was declared by MRI on the basis of an acoustic survey in January 2016 and the HCR from ICES WKICE (2015).

The total landings in the fishing season 2015/2016 amounted to 174 thousand t (preliminary data). One vessel tried capelin fishery in summer 2015 without success, 2500 t were caught in autumn 2015 and the rest during winter months January-March 2015.

The acoustic surveys in autumn 2015 were conducted under difficult weather and ice conditions and the entire distribution area was not covered. The acoustic estimates (6 bill.) were well below 50 bill. The HCR-value that triggers an initial quota. Consequently, ICES advices an initial quota of $0 t$ for the fishing season 2016/17.

## Offshore West Greenland Cod

From 2015 the advice for cod in Greenland offshore waters has been split in two stock components (advice year 2016). The West Greenland offshore stock component is now comprised of the NAFO subdivisions 1A-E in West Greenland. The East Greenland stock component is comprised of the area NAFO subdivision 1F in South Greenland and ICES Subarea 14 in East Greenland.

Some mixing occurs between the two stocks in West Greenland which at present is considered to act as a nursing area for juveniles of the East Greenland stock component. A TAC of 5000 t was introduced in 2015 and 4860 t was fished which is the first time in 25 years that a considerable fishery has taken place. The 2009 YC dominated the catches.

Both the German and Greenland survey indices show that the biomass and abundance has increased due to the 2009 YC, which is present in considerable numbers and a 2010 YC, which is dominating the survey especially in the southern part of the survey area.

The spatial distribution of the 2009 YC is different from previous year classes that usually migrate out of the area at age 4, but a large part of the 2009 YC still remains in the southern area (NAFO 1E) at age 6 in 2015.

No analytical assessment was conducted and there are no biological reference points for the stock. Information from survey indices (German Groundfish survey and Greenland Shrimp and Fish survey) are used as basis for advice. No significant spawning has been observed in the area.

## Inshore Greenland cod

Total catches from the inshore fishery were 25272 t in 2015 which is the highest since early 1990s. Several year classes were caught in the fishery but catches were dominated by the 2010 YC ( 5 yr old).
Survey recruitment indices from the inshore area show that incoming year classes (2012 and 2013) are below average.
The same advice as last year was given as the survey index suggest that the adult part of the stock has fluctuated without trend in recent years, but a relatively high level.

## Cod in East Greenland, South Greenland

From 2014 the management for cod in Greenland offshore waters has been split in two stock components according to areas: NAFO subdivisions 1A-E in West Greenland and NAFO subdivision 1F in South Greenland combined with ICES Subarea 14 in East Greenland. The ICES advice for 2016 has for the first time been given according to these two areas.

The offshore fishery in East and South Greenland in 2015 was conducted as an experimental fishery with a TAC of 18000 tons. Total catches were 15755 tons. The year class dominating the catches was the 2009 YC in Southwest Greenland and the 2008 YC in East Greenland. The largest cod (mean length of 78 cm ) were caught by trawlers on Dohrn Bank close to the Iceland EEZ.

Available survey biomass indices from the Greenland and German surveys show a decline in biomass but the magnitude of decline differs. The Greenland biomass index declined by $16 \%$ but the German biomass index declined by 76\%. The index for 2014 was, however, record high in the German survey, increasing 94\% compared to 2013).

In both surveys the 2009 YC has been dominating in 2013 and 2014 in Southwest Greenland. In 2015 the YC was still dominating the Greenland survey and was still found in Southwest Greenland in large numbers. In the German survey the YC was found in small numbers in 2015 and in equal amount in both South and East Greenland.

Advice was based on an $\mathrm{F}_{\text {proxy }}$ multiplier generated from the relationship between the catches and Greenland survey index in a period with a considered sustainable fishery, multiplied by the latest year's smoothed survey index.

## Greenland Halibut in Subareas 5, 6, 12, and 14

Catches of Greenland halibut in Subareas 5,6,12 and 14 have ranged between 20 and 30 kt in the last two decades and amounts to 25 kt in 2015. The biomass indices used as input to the assessment (combined survey index at Greenland and Iceland) and logbook information from Iceland trawler fishery all show a slight increase in 2015.

A logistic production model in a Bayesian framework has been used to assess stock status and for making predictions. The model includes an extended catch series going
back to the assumed virgin status of the stock at the beginning of the fishery in 1961. Estimated stock biomass showed an overall decline along with the high catches in the late 1980s and early 1990s. Since 2004/2005 the stock has increased slowly and is now at $71 \%$ of Bmsy. Fishing mortality has since 2013 been around $\mathrm{Fmsy}_{\text {a }}$ and is in $201510 \%$ above Fmš. The remaining available indices that are not currently used in the analytical assessment, i.e. logbook from East Greenland trawl fishery and from Faroese trawl fishery and a Faroese survey suggest high biomass in recent years, and therefore supports the recent trend in the assessment. An alternative assessment model (Gadget) was presented in the working group, suggesting similar trends and advice, although slightly more pessimistic about recent trends.

## Golden redfish (Sebastes norvegicus) in Subareas 5, 6 and 14

Total landings in 2015 were 51645 t , which is about 900 t more than in 2014. About 95\% of the catches were taken in Division 5.a. A substantial increase in landings from 14.b since 2010, the highest since early 1990s, and is in relation to a re-established redfish fishery in 2010. Very little redfish is now taken in 5.b.

Catch-at-age data from 5 .a show that the catch was dominated by two strong year classes from 1985 and 1990. From 2008-2011 year classes 1996-1999 were the most important in the fisheries. Their share has reduced relatively fast and the 2000-2005 year classes are now most important contributing about 75\% of the total catch.

Recruitment seems to be low in all areas, both according to the Icelandic groundfish surveys, and the German survey and the Greenland shrimp and fish survey in East Greenland.

The management plan is based on $\mathrm{F}_{9-19}=0.097$ reducing linearly if the spawning stock is estimated below 220000 t ( $\mathrm{B}_{\text {trigger }}$ ). Blim was proposed as 160000 t , lowest SSB in the 2012 run. The 2016 SSB was estimated at $354800 t$, and according to the management plan the TAC advice for 2017 will be 52800 t .

## Icelandic slope (Sebastes mente/la) in 5.a and 14

Total landings of demersal S. mentella in Icelandic waters in 2015 were about 9300 t , $200 t$ less than in 2014. No analytical assessment was conducted and there are no biological reference points for this stock. Survey indices from the Icelandic autumn survey since 2000 are used as basis for advice.

Survey biomass indices show that in Division 5.a the biomass has gradually decreased from 2006-2013, but increased in 2014 and 2015.

The East Greenland shelf is most likely a nursery area for the stock. No new recruits $(<18 \mathrm{~cm})$ are seen in the survey catches of the German survey and the Greenland survey conducted in the area.

Icelandic slope S. mentella is considered a data limited stock (DLS) and follows the ICES framework for such (Category 3.2). When the precautionary approach is applied, catches in 2017 should be no more than 12922 t. All catch are assumed to be landed.

## Shallow Pelagic Sebastes mentel/a

Total landings of shallow pelagic S. mentella in 2015 were 5595 t , a decrease of about 1 000 t compared to 2014. The catches were taken in ICES Subarea 12 and NAFO Division 1 F .

No analytical assessment was conducted and there are no biological reference points for this stock. Survey indices from the international acoustic redfish survey conducted in the Irminger Sea and adjacent waters since 1991 are used as basis for advice.
The last international redfish survey was conducted in June/July 2015, but it did not cover the shallow pelagic stock. The last biomass estimates are from 2013. The results of the acoustic survey show a drastic decrease from 2.2 million t in 1994 to 91000 t in 2013. The next survey is expected to be conducted in June/July 2018.

No signs of recruitment have been observed in the latest German and Greenland surveys on the East Greenland shelf.

## Deep Pelagic Sebastes mentel/a

Total landings of deep pelagic S. mentella s in 2015 were 27433 t , which is 4000 t more than in 2014.
No analytical assessment was conducted and there are no biological reference points for this stock. Survey indices from the biennial international trawl-acoustic redfish survey conducted in the Irminger Sea and adjacent waters since 1999 are used as basis for advice.

The survey was conducted in June/July 2015. A total biomass of 196000 t was estimated, $43 \%$ less than in 2013 ( 280000 t ). Trawl survey estimates in 2013 and 2015 are lowest since the survey started in 1999. The next international trawl-acoustic redfish survey in the Irminger Sea is expected to be conducted in June/July 2018.
No signs of recruitment have been observed in the latest German and Greenland surveys on the East Greenland shelf. Advice for this stock will be given in autumn 2016 following a workshop (WKDEEPRED).

## Greenlandic slope Sebastes mentella in 14.b

In the decade before 2009 S. mentella was mainly a valuable bycatch in the fishery for Greenland halibut. However, since 2009 a fishery directed towards demersal redfish has taken place. Total landings of demersal S. mentella in East Greenland waters in 2015 were 5977 tons, which, except for 2014, is the lowest value the last five years. The proportion of $S$. mentella in the mixed-stock fishery is declining compared to earlier years.

The advice is based on the DLS approach (3.2) using the Greenland survey as basis for advice. Sebastes mentella is a slow growing, late maturing species and is therefore considered vulnerable to overexploitation. Biomass and abundance index for both adult and juvenile redfish have together with the cpue been declining since 2010. Furthermore, on the Greenlandic slope, S. mentella is primarily distributed in a relatively small area, which is also the hot spot area for the fishery. Given the biology of the species, the nature of the fishery and the decline in all stock indicators the advice has to be conservative. Due to the dynamics of the stock and coincident decrease in all stock indicators, the indicator ratio ( $50 \%$ decline) was used and no other uncertainty parameters applied. The advice for 2017 is 1120 tonnes.

### 1.1 Terms of Reference (ToR)

### 1.1.1 Specific ToR

The North-Western Working Group (NWWG), chaired by Rasmus Hedeholm, Greenland, will meet at ICES Headquarters, 27 April-4 May, 2016 to:
a) Address generic ToRs for Regional and Species Working Groups (see below).
b) Answer the special requests from NEAFC: " Pursuant to the Interim guidelines on management of deep sea species [as defined by NEAFC], adopted at the 2014 NEAFC Annual Meeting, ICES is requested to evaluate the provisional categorization of deep sea species adopted by PECMAS and provide necessary clarification on issues highlighted in the document."
c) Prepare for the upcoming workshop of deep pelagic redfish in the Irminger Sea (WKDEEPRED); discuss the approach to be taken by WKDEEPRED to address the ToRs including preliminary analyses/assessments, and additional improvements required in advance of WKDEEPRED.

The assessments will be carried out on the basis of the stock annex. The assessments must be available for audit on the first day of the meeting.
Material and data relevant to the meeting must be available to the group no later than 5 April 2016 according to the Data Call 2016.

For capelin in Iceland-East Greenland-Jan Mayen area, NWWG will agree any changes to the WG type report and the draft advice no later than 10 May 2016.
The assessments will be carried out on the basis of the stock annex in National Laboratories, prior to the meeting.
For capelin in Iceland-East Greenland-Jan Mayen area, Iceland will provide a WG type report and a draft advice sheet on 4 May 2016. NWWG will agree any changes to the WG type report and the Advice sheet no later than 10 May. An ADG will work 24-27 May 2016, and the Advice Release date 10 June 2016.
Other material and data relevant to the meeting must be available to the group no later than 14 days prior to the starting date.
NWWG will report by 10 May 2016 for the attention of ACOM.

| FISH <br> Stock | Stock Name | Stоск <br> Coord. | Assess. <br> Coord. 1 | Assess. <br> Coord. 2 | Advice |
| :---: | :---: | :---: | :---: | :---: | :---: |
| cod-farp | Cod in Subdivision Vb2 (Faroe <br> Bank) | Faroe Islands | Faroe Islands | Faroe Islands | Update |
| cod-farb | Cod in Subdivision Vb2 (Faroe Bank) | Faroe Islands | Faroe Islands | Faroe Islands | Update |
| had-faro | Haddock in Division Vb | Faroe Islands | Faroe Islands | Faroe Islands | Update |
| sai-faro | Saithe in Division Vb | Faroe Islands | Faroe Islands | Faroe Islands | Update |
| cod-iceg | Cod in Division Va (Icelandic cod) | Iceland | Iceland | Iceland | Update |
| had-iceg | Haddock in Division Va (Icelandic haddock) | Iceland | Iceland | Iceland | Update |
| sai-icel | Saithe in Division Va (Icelandic saithe) | Iceland | Iceland | Iceland | Update |
| her-vasu | Herring in Division Va (Icelandic summer-spawners) | Iceland | Iceland | Iceland | Update |
| cap-icel | Capelin in Subareas V, XIV and Division Ila west of $5^{\circ} \mathrm{W}$ (Iceland-East Greenland-Jan Mayen area | Iceland | Iceland | Iceland | Update |
| cod-ingr | Cod (Gadus morhua) in NAFO Subarea 1, inshore (Inshore West Greenland) | Greenland | Greenland | Greenland | Update |
| cod-segr | Cod (Gadus morhua) in ICES Subarea XIV and NAFO Subdivision 1F (East Greenland, South Greenland) | Greenland | Greenland | Germany | Update |
| cod-wgr | Cod (Gadus morhua) in NAFO Subdivision 1A-E (Offshore West Greenland) | Greenland | Greenland | Germany | Update |
| ghl-grn | Greenland halibut in Subareas V, VI, XII and XIV | Greenland | Greenland | Iceland | Update |
| smr-5614 | Redfish (Sebastes marinus) in Subareas V, VI, XII and XIV | Iceland | Iceland | Faroe Islands | Update |
| smn-con | Beaked redfish (Sebastes mentella) in Division Va and Subarea XIV (Icelandic slope stock). | Iceland | Iceland | Germany | Update |
| smn-sp | Beaked Redfish (Sebastes mentella) in Subareas V, XII, XIV and NAFO Subareas $1+2$ (Shallow Pelagic stock < 500 m deep) | Iceland | Germany | Spain | Update |
| smn-dp | Beaked Redfish (Sebastes mentella) in Subareas V, XII, XIV and NAFO Subareas 1+2 (Deep Pelagic stock > 500 m deep) | Iceland | Germany | Spain | Update |
| smn-grl | Beaked Redfish (Sebastes mentella) in Subarea XIV (East Greenland Slope) | Greenland | Greenland | Germany | Update |

### 1.1.2 Generic ToRs for Regional and Species Working Groups

## The working group should focus on:

a) Consider and comment on ecosystem overviews where available;
b) For the fisheries relevant to the working group consider and comment on:
i) descriptions of ecosystem impacts of fisheries where available
ii) descriptions of developments and recent changes to the fisheries
iii) mixed fisheries overview, and
iv) emerging issues of relevance for the management of the fisheries;
c) Conduct an assessment to update advice on the stock(s) using the method (analytical, forecast or trends indicators) as described in the stock annex and produce a brief report of the work carried out regarding the stock, summarizing where the item is relevant:
i) Input data (including information from the fishing industry and NGO that is pertinent to the assessments and projections);
ii) Where misreporting of catches is significant, provide qualitative and where possible quantitative information and describe the methods used to obtain the information;
iii) For relevant stocks estimate the percentage of the total catch that has been taken in the NEAFC Regulatory Area by year in the recent three years.
iv) The developments in spawning-stock biomass, total-stock biomass, fishing mortality, catches (wanted and unwanted landings and discards) using the method described in the stock annex;
v) The state of the stocks against relevant reference points;
vi) Catch options for next year;
vii) Historical performance of the assessment and catch options and brief description of quality issues with these;
d) Produce a first draft of the advice on the fish stocks and fisheries under considerations according to ACOM guidelines.
e) With reference to the Frequency of Assessment criteria agreed by ACOM (see section 5.1 of WGCHAIRS document 03): (1) Complete the calculation of the first set of criteria, by calculating Mohn's rho index for the final assessment year F; (2) Comment on the list of stocks initially identified as candidates for less frequent assessment from the first set of criteria (adding stocks to the list or removing them would require a sufficient rationale to be provided).
f) Estimate precautionary reference points for all the category 1 stocks with undefined PA reference points, following the Technical Guidelines document on reference points developed by ACOM and the WKMSYREF4 report.

## The working group is furthermore requested to:

g) Consider and propose stocks to be benchmarked;
h) Review progress on benchmark processes of relevance to the expert group;
i) Propose specific actions to be taken to improve the quality and transmission of the data (including improvements in data collection);
j) Prepare the data calls for the next year update assessment and for the planned data evaluation workshops;
k) Update, quality check and report relevant data for the stock:
i) Load fisheries data on effort and catches into the INTERCATCH database by fisheries/fleets;
ii) Abundance survey results;
iii) Environmental drivers.

1) Produce an overview of the sampling activities on a national basis based on the INTERCATCH database or, where relevant, the regional database.
$\mathrm{m})$ Identify research needs of relevance for the expert group.

### 1.2 NWWG 2016 work in relation to the generic ToR

The ToRs were not addressed systematically for all stocks. The main focus was on the adoption of assessments that were the basis for stock status and the premise for the forecasts. This was done to ensure that the basis for the advice was agreed upon. Individual report stock sections were not reviewed in plenary due to time constraints, but relevant issues were discussed in plenary and each section was reviewed externally. The summary sheets were all reviewed and agreed upon in plenary.
ad a) Ecosystem overviews were available for the Faroe Islands, Iceland and Greenland ecoregions. In the Icelandic ecosystem the increased temperature/salinity since mid 1990s is regarded as a major factor, which has shifted the distribution of many fish species northwards. The biomass of capelin was previously used directly in the assessment (predicting individual weights of cod). This relationship became less clear in recent years and the weights of cod were therefore estimated in other ways. The assessment of capelin uses an escapement biomass stragtegy, including estimating the removal of capelin biomass by predators. No new ecosystem driver was proposed this year. It was, however, remarked that the effects of important ecosystem drivers was expected to be expressed in the input data of the stock assessments and therefore taken into account in an indirect way. In the Greenland ecoregion the effect of temperature and wind is outlined since these measures are good indicators of the recruitment of cod. These measures are, however, not used directly in the assessments or the advice. In the Faroe ecoregion there has been shown a positive relationship between primary production and the production of demersal fish (cod, haddock and saithe). Primary production is, however, not used directly in the assessments or advice.
ad b) In the overview sections there is a description of the fisheries, including mixed fisheries. In the Icelandic ecoregion the ecosystem effect of fisheries is briefly mentioned with corals being destroyed by fishing gears. Other than this, ecosystem effects of the fisheries were not considered. For Greenland cod both inshore and offshore, the mixing of stocks and apparent changes in distribution was discussed as it is relevant to managers.
ad c-d) All stocks were assessed according to the stock annexes, including forecasts, catch options and retrospective analyses were relevant. This is briefly summarized for all stocks in the report. Only the pelagic redfish stocks are to some extent caught in the

NEAFC regulatory area, and this was quantified. Accordingly, advice drafts were produced for all stocks.
ad e) The group discussed biennial assessment in general under the ICES MSY advice. Here, the derived Fmsy_depends on the assumption of assessment errors. The current values used are probably in all cases based on annual assessment errors. If advice were to be set multi-annually, the input assessment error would need to be set significantly higher when estimating of the advisory Fmş, particularly in stocks were no direct measurements are available of future recruits. The effect of setting higher CV's to accommodate the assessment error would be most pronounced in stocks where the derived Fmsy_is bounded by precautionary consideration (low probability of going below $^{\text {b }}$ $\mathrm{B}_{\mathrm{lim}}$ ). The net effect is that a multi-annual advice should in most cases be based on a lower advisory fishing mortality than used if the advice were annual.
ad f) Precautionary reference points were defined for all relevant stocks according to the guidelines.
ad g-j) No new stocks proposed for a benchmark in 2018. Hence, the Faroese stocks of cod, saithe and haddock are still on schedule to be benchmarked in 2017. The issue lists for these stocks were discussed in plenary as was the data evaluation workshop agenda. The 2015 data call was evaluated for all stocks and updated to facilitate the 2016 data call. There were few issues with the 2015 data call and these were of minor importance to the assessment.
ad k-l) INTERCATCH is not used extensively for any stocks considered in this WG as they are mostly national non-EU stocks, and no data were loaded into INTERCATCH that was not already uploaded. The members of the group were however encouraged to use it and to provide the status on the use of INTERCATCH.
ad $m$ ) These needs as they relate to the assessment procedure were discussed in plenary and summarized in the report.

### 1.3 NWWG 2016 work in relation to the specific ToR

ad a) see above
ad b) The NWWG evaluated the provisional categorization of NEAFC of small redfish (Sebastes viviparus) and Greenland halibut (Reinhardtius hippoglossoides). In both cases the NWWG confirmed the categorization made by NEAFC.
ad c) A subgroup discussed the progress being made and the outstanding issues to address before WKDEEPRED. The co-chair of WKDEEPRED attended as did persons from almost all intuitions expected to participate the working group. Several approaches to generate advice were proposed by WKDEEPRED participants, and all were encouraged to develop these further. A video conference was scheduled in late June, at which point the assessment methods and other material relevant according to the ToR should be presented.

### 1.4 Assessment methods applied to NWWG stocks

The methods applied to assess the stock status of the NWWG stocks covers a wide range from descriptive to age based analytical assessments as follows:

| STOCK | ASSESSMENT MODEL | INPUT* |
| :--- | :--- | :--- |
| Faroe Bank cod | Trend based assessment | Survey |
| Faroe Plateau cod | XSA | Survey |
| Faroe haddock | XSA | Survey |
| Faroe saithe | XSA | cpue** |
| Iceland saithe | ADCAM (statistical catch-at-age) | Survey |
| Iceland cod | ADCAM (statistical catch-at-age) | Survey |
| Iceland haddock | Adapt type model | Survey |
| Iceland herring | NFT-Adapt | Survey |
| Capelin | Linear regression | Survey |
| Inshore West Greenland cod | DLS category 3.2 | Survey |
| East Greenland, South <br> Greenland cod | Fproxy multiplier/ DLS category 3.2 | Survey |
| Offshore West Greenland cod | Descriptive | Survey |
| Greenland halibut | Stock production model (Bayesian) | Survey + cpue |
| S. norvegicus | GADGET (age-length based cohort | Survey |
| S. mentella Iceland slope | Descriptive |  |
| Deep pelagic S. mentella | Descriptive | Survey |
| Shallow pelagic S. mentella | Descriptive | Survey + cpue |
| S. mentella Greenland Slope | Descriptive | Survey + cpue |

* Landings or landings by age are input to all assessments
** The cpue is adjusted by survey information about distribution width.


### 1.5 Benchmarks and workshops

No stocks were benchmarked in the preceding year. Faroe cod (plateau and bank), Faroe saithe and Faroe haddock are scheduled for a 2017 benchmark. Faroe saithe was added to the list of stocks to benchmark in 2017 along with the other Faroese stocks in 2017.

### 1.6 Chair

This is the second year for Chair, Rasmus Hedeholm, Greenland, who is scheduled to chair the group from 2015-2017.

## 2 Demersal Stocks in the Faroe Area (Division 5b and Subdivision 2a4)

### 2.1 Overview

### 2.1.1 Fisheries

The main fisheries in Faroese waters are mixed-species, demersal fisheries and single species pelagic fisheries. The demersal fisheries are mainly conducted by Faroese vessels, whereas the pelagic fisheries are conducted both by Faroese vessels and by foreign vessels licensed through bilateral and multilateral fisheries agreements. The usual picture changed in 2011, however, since no mutual agreement could be reached between the Faroe Islands and the EU and Norway, respectively, due to the dispute regarding the share of mackerel. From 2013, the agreement has been re-established.

Pelagic Fisheries. Three main species of pelagic fish are fished in Faroese waters: blue whiting, herring and mackerel; several nations participate. The Faroese pelagic fisheries are conducted by purse-seiners, larger purse-seiners also equipped for pelagic trawling and trawlers otherwise performing demersal fisheries. The pelagic fishery by Russian vessels is conducted by large factory trawlers. Other countries use purseseiners and factory trawlers.

Demersal Fisheries. Although they are conducted by a variety of vessels, the demersal fisheries can be grouped into fleets of vessels operating in a similar manner. Some vessels change between longlining, jigging and trawling, and they therefore can appear in different fleets. The number of licenses can be found in Table 2.3. The grouping of the vessels under the management scheme can be seen in section 2.1.3.

### 2.1.2 Fisheries and management measures

The fishery around the Faroe Islands has for centuries been an almost free international fishery involving several countries. Apart from a local fishery with small wooden boats, the Faroese offshore fishery started in the late 19th century. The Faroese fleet had to compete with other fleets, especially from the UK with the result that a large part of the Faroese fishing fleet became specialized in fishing in other areas. So except for a small local fleet most of the Faroese fleet were fishing around Iceland, at Rockall, in the North Sea and in more distant waters like the Grand Bank, Flemish Cap, Greenland, the Barents Sea and Svalbard.

Up to 1959, all vessels were allowed to fish around the Faroes outside the 3 nm zone. During the 1960s, the fisheries zone was gradually expanded, and in 1977 an EEZ of 200 nm was introduced in the Faroe area. The demersal fishery by foreign nations has since decreased and Faroese vessels now take most of the catches. The fishery may be considered a multifleet and multispecies fishery as described below.

During the 1980s and 1990s the Faroese authorities have regulated the fishery and the investment in fishing vessels. In 1987 a system of fishing licenses was introduced. The demersal fishery at the Faroe Islands has been regulated by technical measures (minimum mesh sizes and closed areas). In order to protect juveniles and young fish, fishing is temporarily prohibited in areas where the number of small cod, haddock and saithe exceeds $30 \%$ (in numbers) of the catches; after 1-2 weeks, sometimes longer, the areas are again opened for fishing. A reduction of effort has been attempted through banning of new licenses and buy-back of old licenses.

A quota system, based on individual quotas, was introduced in 1994. The fishing year started on 1 September and ended on 31 August the following year. The aim of the quota system was, through restrictive TACs for the period 1994-1998, to increase the SSBs of Faroe Plateau cod and haddock to 52000 t and 40000 t , respectively. The TAC for saithe was set higher than recommended scientifically. It should be noted that especially cod and haddock but also saithe are caught in a mixed fishery and any management measure should account for this. Species under the quota system were Faroe Plateau cod, haddock, saithe, redfish and Faroe Bank cod.

The catch quota management system introduced in the Faroese fisheries in 1994 was met with considerable criticism and resulted in discarding and in misreporting of portions of the catches. Reorganization of enforcement and control did not solve the problems. As a result of the dissatisfaction with the catch quota management system, the Faroese Parliament discontinued the system as from 31 May 1996. In close cooperation with the fishing industry, the Faroese government developed a new system based on individual transferable effort quotas in days within fleet categories. The new system entered into force on 1 June 1996. The fishing year from 1 September to 31 August, as introduced under the catch quota system, has been maintained.

The individual transferable effort quotas apply to 1) the longliners less than 110 GRT, the jiggers, and the single trawlers less than 400 HP (Groups 4,5), 2) the pairtrawlers (Group 2) and 3) the longliners greater than 110 GRT (Group 3). The single trawlers greater than 400 HP were in 2011 included into the fishing days system and were allocated a number of fishing days (Tables 1 and 2). They are not allowed to fish within the 12 nautical mile limit and the areas closed to them, as well as to the pairtrawlers, have increased in area and time. Their catch of cod and haddock was before 2011 limited by maximum bycatch allocation. This fleet started to pair-trawl, and since the fiscal year 2011/12, merged with the pairtrawlers group. The single trawlers less than 400 HP are given special licenses to target flatfish inside 12 nautical miles with a bycatch allocation of $30 \%$ cod and $10 \%$ haddock. In addition, they are obliged to use sorting devices in their trawls in order to minimize their bycatches. One fishing day by longliners less than 110 GRT is considered equivalent to two fishing days for jiggers in the same gear category. Longliners less than 110 GRT could therefore double their allocation by converting to jigging. Table 2.1 shows the allocated number of fishing days by fleet group since the fiscal year 1996/1997 and in Table 2.2 is a comparison between number of allocated days and number of actually used fishing days. From Table 1 it can been seen that since 1996/1997, the number of days allocated has been reduced considerable and is now $50 \%$ of the originally allocated days. Despite this, there still are many unused days in the system (Table 2.2).

Holders of individual transferable effort quotas who fish outside the thick line on Figure 2.2 can fish for 3 days for each day allocated inside the line. Trawlers are generally not allowed to fish inside the 12 nautical mile limit. Inside the innermost thick line only longliners less than 110 GRT and jiggers less than 110 GRT are allowed to fish. The Faroe Bank shallower than 200 m is closed to trawling. Due to the serious decline of the Faroe Bank cod, the Bank has been closed since 1 January 2009 for all gears except for a minor jigging fishery during summertime.

The fleet segmentation used to regulate the demersal fisheries in the Faroe Islands and the regulations applied are summarized in Table 2.3.

The effort quotas are transferable within gear categories. The allocations of number of fishing days by fleet categories was made such that together with other regulations of the fishery they should result in average fishing mortalities on each of the 3 stocks of
0.45 , corresponding to average annual catches of $33 \%$ of the exploitable stocks in numbers. Built into the system is also an assumption that the day system is self-regulatory, because the fishery will move between stocks according to the relative availability of each of them and no stock will be overexploited. These target fishing mortalities have been evaluated during the 2005 and 2006 NWWG meetings. The realized fishing mortalities have been substantially higher than the target for cod, appear to have been almost at the target for saithe in recent years, while for haddock, fishing mortality remains below the target.

In addition to the number of days allocated in the law, it is also stated in the law what percentage of total catches of cod, haddock, saithe and redfish, each fleet category on average is expected to fish. These percentages are as follows:

| FLEET CATEGORY | COD | HADDOCK | SAITHE | REDFISH |
| :--- | :---: | :---: | :---: | :---: |
| Longliners < 110GRT, |  |  |  |  |
| Jiggers, single trawl. < 400HP | $51 \%$ | $58 \%$ | $17.5 \%$ | $1 \%$ |
| Longliners > 110GRT | $23 \%$ | $28 \%$ |  |  |
| Pairtrawlers | $21 \%$ | $10.25 \%$ | $69 \%$ | $8.5 \%$ |
| Single trawlers > 400 HP | $4 \%$ | $1.75 \%$ | $13 \%$ | $90.5 \%$ |
| Others | $1 \%$ | $2 \%$ | $0.5 \%$ | $0.5 \%$ |

The technical measures as mentioned above are still in effect. An additional measure to reduce the fishing mortality on cod and haddock and to especially reduce the mortality on the youngest age groups was introduced (See the 2013 NWWG report, Figure 2.3) in July 2011, but was terminated in August 2013.

### 2.1.3 The marine environment and potential indicators

The waters around the Faroe Islands are in the upper 500 m dominated by the North Atlantic current, which to the north of the islands meets the East Icelandic current. Clockwise current systems create retention areas on the Faroe Plateau (Faroe shelf) and on the Faroe Bank. In deeper waters to the north and east and in the Faroe Bank channel there is deep Norwegian Seawater, and to the south and west is Atlantic water. From the late 1980s the intensity of the North Atlantic current passing the Faroe area decreased, but it has increased again in the most recent years. The productivity of the Faroese waters was very low in the late 1980s and early 1990s. This applies also to the recruitment of many fish stocks, and the growth of the fish was poor as well. Since then, there have been several periods with high or low productivity, which has been reflected in the fish landings a couple of years afterwards.

There has been observed a clear relationship, from primary production to the higher trophic levels (including fish and seabirds), in the Faroe shelf ecosystem, and all trophic levels seem to respond quickly to variability of primary production in the ecosystem (Gaard, E. et al. 2002). There is a positive relationship between primary production and the cod and haddock individual fish growth and recruitment $1 / 2-2$ years later. The primary production index has been below average since 2002 except for 2004 and 20082010 when it was above average (Figure 2.3). The estimate of primary production in 2016 will not be available until July. The primary production index could therefore be a candidate ecosystem and stock indicator. Another potential indicator candidate is the
so-called Subpolar Gyre Index, which is an index for the primary production in the outer areas (Figure 2.3).

Recent work (Steingrund et al., 2012) shows that there is a moderate positive correlation between primary production on the Faroe Shelf and the subsequent production of cod (Steingrund and Gaard, 2005). There is also a moderate positive correlation for haddock and saithe. If all three species are combined, the positive correlation becomes stronger (Figure 2.4). However, the last period of high productivity (2008-2010) did not lead to any marked increase in the stock size of cod/haddock, but only in saithe. The catchability of cod with longlines also increased by a factor of 2-3 in the same period.

### 2.1.4 Summary of the 2016 assessment of Faroe Plateau cod, haddock and saithe

A summary of selected parameters from the 2016 assessment of Faroe Plateau cod, Faroe haddock and Faroe saithe is shown in Figure 2.6. As mentioned in previous reports of this WG, landings of cod, haddock and saithe on the Faroes appear to be closely linked with the total biomass of the stocks.

For cod, the exploitation ratio and fishing mortality have remained relatively stable over time, although they have been more fluctuating in recent years (Figure 2.6). For haddock, the exploitation rate was high in the 1930s and decreasing from the 1950s and 1960s, while it has been fluctuating since the mid-1970s. For saithe, the exploitation rate was low in the 1930s and 1950s and increased until the 1970s, it decreased from the early 1990s-1998 and has increased close to the highest values observed in 2009. It has since declined again.

Another main feature of the plots of landings, biomasses, mortalities and recruitment is the apparent periodicity during the time-series with cod and haddock showing almost the same fluctuations and time-trends. Moreover, while the sum of cod, haddock and saithe biomasses has been rather constant over time (varied between 300-500 thousand tonnes most years), the proportion of saithe has increased during the period from 1924 up to today whereas the proportion of cod has decreased (Figure 2.6).

### 2.1.5 Reference points for Faroese stocks

As explained elsewhere in this report, MSY reference points were estimated for cod and haddock in 2011 and for saithe in 2014 in addition to the already existing PA reference points. These reference points are all estimated based on single-species models. Multispecies models may give very different perception of $\mathrm{F}_{\mathrm{MSY}}$ reference points than single-species models, and for the Faroe area this could be extra true, since there is a close relationship between the environment and the fish stocks and between fish stocks (see section 2.1.3). For example, adding the recruitment of cod and haddock and relating them to zooplankton concentration shows a strong negative correlation (Figure 2.5), but a potential causal relationship is unknown.

Faroe saithe stock dynamics is puzzling. If the biomass estimates prior to 1961 are approximately correct then there has been an increase in biomass from 1925 up to now as well as in catch and exploitation rate. There might be an interaction with cod, since the cod biomass has decreased over the same period. It might be speculated that trawling activity in the deep areas ( $>150 \mathrm{~m}$ ) from the 1950s has had a negative effect on cod and a positive effect on saithe. Hence, it might not be possible to maximize cod and saithe catches at the same time.

### 2.1.6 Management plan

In 2011 the Faroese minister of fisheries established a group of experts to formulate a management plan for cod, haddock and saithe including a harvest control rule and a recovery plan. The group consisted of scientists from the Faroe Marine Research Institute and the Faroese University, of 1 representative from the industry (trawlers) and 1 from the Ministry of Fisheries. The results of this work was delivered to the Minister of Fisheries in spring 2012 but the outcome has not been approved by the authorities so far and not been implemented. Basically, the plan builds on the MSY framework developed by ICES.

### 2.1.7 Oher issues

In order to put the current assessment into a wider context, the biomass of Faroe saithe was estimated back to 1925 by scaling cpue values for English steam trawlers to the biomass obtained from the stock assessment, see Working Document 13. The cpue series was from 1924-1978 and the stock assessment from 1961-2015. The overlapping years 1961-1971 were used as a basis to scale the saithe biomass back to 1924. Since the biomass estimates were rather noisy, a three-year moving average was taken as the final estimate of biomass back in time (Table 2.1.7.1). The table shows that the saithe biomass prior to 1960, when there was little fishery for saithe, was lower than during the fishery intensive period after 1970.

### 2.1.8 References:

Gaard. E., Hansen, B., Olsen, B and Reinert, J. 2001. Ecological features and recent trends in physical environment, plankton, fish stocks and seabirds in the Faroe plateau ecosystem. In: KSherman and H-R Skjoldal (eds). Changing states of the Large Marine Ecosystems of the North Atlantic.

Steingrund, P., and Gaard, E. 2005. Relationship between phytoplankton production and cod production on the Faroe Shelf. ICES Journal of Marine Science, 62: 163-176.

Steingrund, P., and Hátún, H. 2008. Relationship between the North Atlantic Subpolar Gyre and fluctuations of the saithe stock in Faroese waters. NWWG 2008 Working Document 20.
Steingrund, P., Gaard, E., Reinert, J., Olsen, B., Homrum, E., and Eliassen, K. 2012. Trophic relationships on the Faroe Shelf ecosystem and potential ecosystem states. In: Homrum, E., 2012. The effects of climate and ocean currents on Faroe Saithe. PhD-thesis, 2012.

Table 2.1. Number of allocated days since the fiscal year 1996/97.

| Allocated number of days: |  |  |  |  |  |  |  |  |  |  |  |  | Available |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bólkur | $\begin{array}{\|l} \hline \text { Smb. LI.: } \\ (5020 / 5-96) \\ \hline \end{array}$ | Serlig viठ̃m.(12/15 mdr!) | 1 ytri | 1 innaru | 2 ytri | 2 innari | 3 | 4 A | 4 B | 4 D | 4 T | 5 | (at ráda yvir) | Dagar tils. |
| 1996/97 |  |  |  |  |  | 8225 | 3040 | 4700 | 3080 | 1540 |  | 22000 | 1000 | 43585 |
| 1996/97 | $\begin{aligned} & (846 / 6-97) \\ & (1339 / 8-97) \end{aligned}$ | $\begin{gathered} \text { (12/15mdr!) } \\ 12 \mathrm{mdr}! \end{gathered}$ |  |  |  | 8225 | 3040 | 5600 | 3410 | 1650 |  | 27000 | 660 | 49585 |
| 1997/98 |  |  |  |  |  | 7199 | 2660 | 4696 | 4632 |  |  | 23625 | 577 | 43389 |
| 1998/99 |  |  |  |  |  | 6839 | 2527 | 4461 | 4400 |  |  | 22444 | 548 | 41219 |
| 1999/2000 |  |  |  |  |  | 6839 | 2527 | 4461 | 4400 |  |  | 22444 | 548 | 41219 |
| 2000/2001 | $\left\{\begin{array}{l} \text { (104 17/8-00) } \\ (11515 / 8-01) \\ (7613 / 8-02) \end{array}\right.$ |  |  |  |  | 6839 | 2527 | 4461 | 4400 |  |  | 22,444 | 548 | 41219 |
| 2001/2002 |  |  |  |  |  | 6839 | 2527 | 4461 | 4400 |  |  | 22444 | 0 | 40671 |
| 2002/2003 |  |  |  |  |  | 6771 | 2502 | 4416 | 4356 |  |  | 22220 | 0 | 40265 |
| 2003/2004 | $\left\{\begin{array}{l} (1008 / 8-03) \\ (49 ~ 18 / 8-04) \end{array}\right.$ |  |  |  |  | 6636 | 2452 | 4328 | 4269 |  |  | 21776 | 0 | 39461 |
| 2004/2005 |  |  |  |  |  | 6536 | 2415 | 4263 | 4205 |  |  | 21449 | 0 | 38868 |
| 2005/2006 | $\left(\begin{array}{l} (98 ~ 19 / 8-05) \\ (81 \\ 17 / 8-06) \end{array}\right.$ |  |  |  |  | 5752 | 3578 | 1770 | 2067 |  | 1766 | 21235 | 0 | 36168 |
| 2006/2007 |  |  |  |  |  | 5752 | 3471 | 1717 | 2005 |  | 1713 | 20598 | 0 | 35256 |
| 2007/2008 | (8020/8-07) |  |  |  |  | 5637 | 3402 | 1683 | 1965 |  | 1679 | 20186 | 0 | 34552 |
| 2008/2009 | (76 15/8-08) |  |  |  |  | 5073 | 3062 | 1515 | 1769 |  | 1511 | 18167 | 0 | 31097 |
| 2008/2009 | (62 25/5-09) |  |  |  |  | 4638 | 3095 | 1393 | 1848 |  | 1621 | 18167 | 0 | 30762 |
| 2009/2010 | (106 17/8-09 |  |  |  |  | 4406 | 2940 | 1323 | 1756 |  | 1540 | 17259 | 0 | 29224 |
| 2010/2011 | (87 18/8-10) |  | 1700 | 900 |  | 4274 | 2852 | 1323 | 1756 |  | 1540 | 13259 | 0 | 25004 |
| 2010/2011 | sama - |  | 1700 | 900 |  | 4274 | 2852 | 1323 | 1756 |  | 1540 | 13259 | 0 | 27604 |
| 2011/12 | $\begin{array}{\|l\|} \hline(105 ~ 18 / 8-11) \\ (112 \\ 2 / 9-11) \end{array}$ |  |  |  | 1530 | 4657 | 2567 | 1058 | 1405 |  | 1386 | 10607 |  | 23210 |
| 2012/13 | (89 17/8-12) |  |  |  | 1530 | 4626 | 2567 | 1011 | 1533 |  | 1386 | 10607 |  | 23260 |
| 2013/14 | (109 16/8-13) |  |  |  | 1530 | 4441 | 2387 | 1011 | 1533 |  | 1386 | 9865 |  | 22153 |
| 2014/15 | (L89-18/8-14) |  |  |  | 1530 | 4455 | 2387 | 1029 | 1530 |  | 1386 | 9865 |  | 22182 |
| 2015/16 | $(\text { L108-5/8-15) }$ |  |  |  | 1530 | 4455 | 2387 | 1029 | 1530 |  | 1386 | 9865 |  | 22182 |

Table 2.2. Number of days allocated and the number actually used since the fiscal year 2014/2015

|  |  |  |  | pr. 10. mars. | 2016 (61/3 m |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fleet segment | Allocated days <br> 2014/15 | Used <br> days <br> pr. Dato | \% used days | $\begin{aligned} & \text { Allocated } \\ & \text { days } \\ & 2015 / 16 \end{aligned}$ | Used days pr. Dato | \% used days |
| Reference: | (L89-18/8-14) |  |  | (L108-5/8-15) |  |  |
| Group 1 - innaru leióir |  |  |  |  |  |  |
| Group 1 - ytri leiõir |  |  |  |  |  |  |
| Group 2 - (innaru leiôi | 4455 | 4,307.87 | 97\% | 4455 | 2000.70 | 45\% |
| Group 2 - ytri leiôir | 1530 | 1,125.41 | 74\% | 1530 | 524.34 | 34\% |
| Group 3 | 2387 | 1234.57 | 52\% | 2387 | 939.92 | 39\% |
| Group 4A | 1029 | 253.59 | 25\% | 1029 | 167.07 | 16\% |
| Group 4B | 1530 | 565.34 | 37\% | 1530 | 424.94 | 28\% |
| Group 4T | 1386 | 716.83 | 52\% | 1386 | 371.1 | 27\% |
| Group 5A | 2640 | 1297 | 49\% | 2310 | 486 | 21\% |
| Group 5B | 7225 | 3709 | 51\% | 7555 | 1697 | 22\% |
| Total | 22182 | 13,209.61 | 60\% | 22182 | 6611.07 | 30\% |


| Estimation of the whole year |  | Tillutaбsmb.Vørn(05/10-15) |
| :---: | :---: | :---: |
| Mett ársnýtsla <br> Faktor <br> Væntandi: |  |  |
|  | 1.895 |  |
|  |  |  |
| (L108-5/8-15) | Predicted | óbroytt |
|  |  | (10/3-16) |
|  |  |  |
| 3,791.00 | 82\% | 4,353.25 |
| 993.54 | 65\% | 1,522.83 |
| 1,780.99 | 69\% | 2,148.22 |
| 316.57 | 31\% | 595.76 |
| 805.19 | 53\% | 932.62 |
| 703.17 | 51\% | 1,180.84 |
| 920.89 | 19\% | 2310 |
| 3,215.54 | 55\% | 7555 |
| 12,526.90 | 54\% | 20,598.52 |

Table 2.3. Main regulatory measures by fleet in the Faroese fisheries in 5 b. The fleet capacity is fixed, based on among other things no. of licenses. Number of licenses within each group (by May 2006) are as follows: 1: 12; 2:29; 3:25; 4A: 25; 4B: 21; 4T: 19; 5A:140; 5B: 453; 6: 8. These licenses have been fixed in 1997, but in group 5B a large number of additional licenses can be issued upon request.

|  | Fleet segment | Subgroups |  | Main regulation tools |
| :---: | :---: | :---: | :---: | :---: |
| 1 | Single trawlers > $400 \mathrm{HP}$ | $\begin{aligned} & \text { non } \\ & \mathrm{e} \end{aligned}$ |  | Fishing days, have from 2011/12 been merged with the pairtrawlers, area closures |
| 2 | $\begin{aligned} & \text { Pairtrawlers > } 400 \\ & \text { HP } \end{aligned}$ | $\begin{aligned} & \text { non } \\ & \mathrm{e} \end{aligned}$ |  | Fishing days, area closures |
| 3 | $\text { Longliners > } 110$ <br> GRT | $\begin{aligned} & \text { non } \\ & \mathrm{e} \end{aligned}$ |  | Fishing days, area closures |
| 4 | Coastal vessels>15 GRT | 4A | Trawlers 15-40 GRT | Fishing days |
|  |  | 4A | Longliners 15-40 GRT | Fishing days |
|  |  | 4B | Longliners>40 GRT | Fishing days |
|  |  | 4 T | Trawlers>40 GRT | Fishing days |
| 5 | Coastal vessels $<15$ GRT | 5A | Full-time fishers | Fishing days |
|  |  | 5B | Part-time fishers | Fishing days |
| 6 | Others |  | Gillnetters | Bycatch limitations, fishing depth, no. of nets |
|  |  |  | Others | Bycatch limitations |



Figure 2.1. The 2015 distribution of fishing activities by some major fleets. From top: Gillnet, longline $>1010 \mathrm{HP}$, trap and trawl. The longline fleet below 110 GRT is not shown here since they are not obliged to keep logbooks.

Exclusion zones for trawling

| Area | Period |
| :---: | :---: |
| a | 1 jan -31 des |
| aa | 1 jun -31 aug |
| b | 20 jan -1 mar |
| c | 1 jan -31 des |
| d | 1 jan -31 des |
| e | 1 apr -31 jan |
| f | 1 jan -31 des |
| g | 1 jan -31 des |
| h | 1 jan -31 des |
| i | 1 jan -31 des |
| j | 1 jan -31 des |
| k | 1 jan -31 des |
| l | 1 jan -31 des |
| m | 1 feb -1 jun |
| n | 31 jan -1 apr |
| o | 1 jan -31 des |
| p | 1 jan -31 des |
| r | 1 jan -31 des |
| s | 1 jan -31 des |
| C 1 | 1 jan -31 des |
| C 2 | 1 jan -31 des |
| C 3 | 1 jan -31 des |

Spawning closures

| Area | Period |
| :---: | :---: |
| 1 | 15 feb -31 mar |
| 2 | 15 feb -15 apr |
| 3 | 15 feb -15 apr |
| 4 | 1 feb -1 apr |
| 5 | 15 jan -15 mai |
| 6 | 15 feb -15 apr |
| 7 | 15 feb -15 apr |
| 8 | 1 mar -1 may |

Figure 2.2. Fishing area regulations in Division 5b. Allocation of fishing days applies to the area inside the outer thick line on the Faroe Plateau. Holders of effort quotas who fish outside this line can triple their numbers of days. Longliners larger than 110 GRT are not allowed to fish inside the inner thick line on the Faroe Plateau. If longliners change from longline to jigging, they can double their number of days. The Faroe Bank shallower than 200 m depths ( $\mathrm{a}, \mathrm{aa}$ ) is regulated separate from the Faroe Plateau. It is closed to trawling and the longline fishery is regulated by individual day quotas.



Figure 2.3. Temporal development of the phytoplankton index over the Faroe Shelf area ( $<\mathbf{1 3 0} \mathbf{m}$ ) and the Subpolar Gyre index which indicates productivity in deeper waters.


Figure 2.4. Relationship between primary production and production of cod, haddock and saithe.


Figure 2.5. Relationship between zooplankton concentration and recruitment of cod and haddock on the Faroe Plateau.


Figure 2.6. Summary of the stock dynamics for Faroe Plateau cod, Faroe haddock and Faroe saithe.

Table 2.1.7.1. Saithe biomass (age 3+) 1925-2015 in tons. Year label is sum of first row and first column.

|  | 1925 | 1950 | 1975 | 2000 |
| :---: | :---: | :---: | :---: | :---: |
| 0 | 86621 | 144660 | 189794 | 227759 |
| 1 | 91614 | 133220 | 181587 | 292429 |
| 2 | 102955 | 154250 | 168856 | 332716 |
| 3 | 121390 | 151612 | 150481 | 330728 |
| 4 | 137929 | 158206 | 127858 | 323589 |
| 5 | 127712 | 157146 | 134902 | 338048 |
| 6 | 115346 | 146198 | 153827 | 269512 |
| 7 | 111440 | 142441 | 165608 | 218366 |
| 8 | 118018 | 127763 | 187681 | 189274 |
| 9 | 123435 | 120344 | 198291 | 153457 |
| 10 | 121033 | 111550 | 192715 | 134874 |
| 11 | 116390 | 105008 | 238071 | 126661 |
| 12 | 119400 | 111502 | 254322 | 115439 |
| 13 | 124442 | 129760 | 261993 | 115660 |
| 14 | 139453 | 139221 | 231424 | 158781 |
| 15 | 150543 | 150302 | 193901 | 194118 |
| 16 | 161634 | 162609 | 151523 |  |
| 17 | 172724 | 161578 | 125365 |  |
| 18 | 183815 | 170304 | 134696 |  |
| 19 | 194905 | 197298 | 128717 |  |
| 20 | 205996 | 212238 | 154403 |  |
| 21 | 228522 | 218372 | 163848 |  |
| 22 | 207472 | 234819 | 183162 |  |
| 23 | 193811 | 210057 | 166861 |  |
| 24 | $141145$ | 205557 | 214410 |  |

### 3.1 State of the stock

Total nominal catches of the Faroe Bank cod from 1987 to 2015 as officially reported to ICES are given in Table 3.1 and since 1965 in Figure 3.1. UK catches reported to be taken on the Faroe Bank are all assumed to be taken on the Faroe Plateau and are therefore not used in the assessment. Landings have been highly variable from 1965 to the mid-1980s, reflecting the opportunistic nature of the cod fishery on the Bank, with peak landings slightly exceeding 5000 t in 1973 and 2003. The trend of landings has been smoother since 1987, declining from about 3500t in 1987 to only 330 t in 1992 before increasing to 3600 t in 1997. In 2015 landings were estimated at 17 t which is the lowest ever recorded since 1965 (Figure 3.1). Longline fishing effort increased substantially in 2003 and although it decreased in 2004 and 2005 the latter remains the second highest fishing effort observed since 1988 (Figure 3.1). From 2005-2007 the effort has been reduced substantially. In the 2010/2011 and 2011/2012 fishing years a total of 61 and 100 fishing-days were allocated to the Bank. No days have been allocated since 2012.
The Faroese groundfish surveys (spring and summer) cover the Faroe Bank and cod is mainly taken within the 200 m depth contour. The catches of cod per trawl hour in depths shallower than 200 meter are shown in Figure 3.2.
Spring survey was initiated in 1983 and discontinued in 1996, 2004 and 2005. Summer survey has been carried out since 1996. The cpue of spring survey was low during 1988-1995 varying between 73 and 95 kg per tow. Although noisy, the survey suggests higher, possibly increasing biomass during 1995-2003. Spring index suggests that the stock increased in 2013 and 2014 but it decreased rapidly in 2015 and 2016 well below the average of that of the period 1996-2002. The 2015 summer index is estimated at 25 kg per tow and the 2016 spring survey at 19 kg per tow, which are among the lowest values in both series. There are conflicting signals between both indices from 2012 to 2014. The agreement between summer and spring index is good during 1996 to 2001 and since 2006, but they diverged in the 2002-2003 and 2012-2014 periods. Both indices have remained well below average since 2004.

The figure of length distributions (figures 3.3 and 3.5) show in general good recruitment of 1 year old in summer survey from 2000-2002 (lengths $26-45 \mathrm{~cm}$ ), corresponding to good recruitment of 2 years old in spring surveys from 2001 to $2003(40-60 \mathrm{~cm})$. Spring index shows poor recruitment from 2006-2016 reflecting the weak year classes observed in summer survey since 2004. Age-disaggregated indices confirm the pattern observed in the length composition (figure 3.4 and figure 3.6)
A way to estimate recruitment strength is by simply counting the number of fish in length groups in the surveys. In spring index, recruitment was estimated as total number of fish below 60 cm (2-year old) and in summer index as number of fish below 45 cm (1-year old). According to summer index the recruitment of 1 year old was good from 2000 to 2003, while the recruitment has been relatively poor since 2004 (Figure 3.7) Spring recruitment index in 2015 shows no sign of incoming year classes. Correlation between spring and summer survey recruitment indices is fairly good ( $\mathrm{r}=0.86$ ). Correlation between numbers of 1 -year and 2 -years old cod in the age-disaggregated summer and spring surveys respectively is estimated at $\mathrm{r}=0.79$.

The group tried the ASPIC (Prager 1992) stock production model for the stock. The model requires catch data and corresponding effort or cpue data that are reasonable indices of the stock biomass.

ASPIC requires starting guesses for $r$, the intrinsic rate of increase, MSY, B1/BMSY ratio and $q$, catchability coefficients. No sensitivity analysis was performed to explore the stability of parameter estimation.

The program was run with the time-series from 1983-2015 including spring survey and 1996-2015 summer cpue's separately. The result of the runs are presented in tables 3.2 and 3.3 For both runs the model seemed to follow reasonably well survey trends in periods of low stock abundances but it failed to pick up the large increases observed in the 1996-2003 period (figures 3.8 and 3.9).

However estimates of $r=0.34$ and $\mathrm{Fms}_{\text {M }}=0.17$ (using autumn survey series) seem spurious given that the Faroe Bank cod is the fastest growing cod stock in the Atlantic.

The ratio of landings to the survey indices provides an exploitation ratio, which can be used as a proxy to relative changes in fishing mortality. For summer survey, the results suggest that fishing mortality has been reasonably stable during 1996 to 2002, but that it increased steeply in 2003, consistent with the $160 \%$ increase in longline fishing days in that year (Figure 3.1). The exploitation ratio has decreased since 2006 but increased in 2011 due to the increase in catches and decreased again afterwards reflecting autumn of catches observed since 2011.

### 3.2 Comparison with previous assessment and forecast

The status of the stock remains almost unchanged with respect to last year's assessment. Both spring and summer indices suggest the stock is well below average while there are no indications of incoming recruitment. Spring index suggests an increasing stock biomass from 2012-2014 which it is however not picked up by summer survey. The exploratory production model performed since 2013 confirms the poor status of the stock.

### 3.3 Management plans and evaluations

None

### 3.4 Management considerations

The landing estimates are uncertain because since 1996 vessels are allowed to fish both on the Plateau and on Faroe Bank during the same trip, rendering landings from both areas uncertain. Given the relative size of the two fisheries, this is a bigger problem for Faroe Bank cod than for Faroe Plateau cod, but the magnitude remains unquantified for both. The ability to provide advice depends on the reliability of input data. If the cod landings from Faroe Bank are not known, it is difficult to provide advice. If the fishery management agency intends to manage the two fisheries to protect the productive capacity of each individual unit, then it is necessary to identify the catch removed from each stock. Simple measures should make it possible to identify if the catch is originating in the Bank or from the Plateau e.g. by storing in different section of the hold and/or by tagging of the different boxes.

Consistent with the advice given in 2015 the WG suggests the closure of the fishery until the recovery of the stock is confirmed. The reopening of the fishery should not be considered until both surveys indicate a biomass at or above the average that of the period 1996-2002.

### 3.5 Regulations and their effects

In 1990, the decreasing trends in cod landings from Faroe Bank lead ACFM to advise the Faroese authorities to close the bank to all fishing. This advice was followed for depths shallower than 200 meters. In 1992 and 1993 longliners and jiggers were allowed to participate in an experimental fishery inside the 200 meters depth contour. For the quota year 1 September 1995 to 31 August 1996 a fixed quota of 1050 t was set. The new management regime with fishing days was introduced on 1 June 1996 allowing longliners and jiggers to fish inside the 200 m contour. The trawlers are allowed to fish outside the 200 m contour.

A total fishing ban during the spawning period (1 March-1 May) has been enforced since 2005. In 2009, fishing was restricted to all fishing gears from 1 January-31 August. However, in the 2010/2011 and 2011/2012 fishing years a total of 61 and 100 fishingdays were allocated to the Bank to jiggers in the shallow waters of the Bank. No days have been allocated since 2012.

Table 3.1. Faroe Bank (subdivision Vb2) cod. Nominal catches (tonnes) by countries 1986-2015 as officially reported to ICES. From 1992 the catches by Faroe Islands and Norway are used in the assessment.

|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 1836 | 3409 | 2966 | 1270 | 289 | 297 | 122 | 264 | 717 | 561 | 2051 | 3459 | 3092 | 1001 |  |  |
| Norway | 6 | 23 | 94 | 128 | 72 | 38 | 32 | 2 | 8 | 40 | 55 | 135 | 147 | 88 |  |  |
| UK (EW/N) | - | - | - | - | $2^{\text {F/ }}$ | $1{ }^{2}$ | $74{ }^{5}$ | $186^{\text {² }}$ | $56^{\text {² }}$ | $43^{5}$ | $126^{5}$ | $61^{5}$ | $27^{5}$ | - |  |  |
| UK (Scotland) | $63{ }^{\text {5 }}$ | $47^{5}$ | $37^{\text {5 }}$ | $14^{5}$ | $205^{5}$ | $90^{5}$ | $176{ }^{5}$ | $118{ }^{\text {/ }}$ | $227{ }^{\text {/ }}$ | $551{ }^{5}$ | $382^{\frac{5}{3}}$ | $277{ }^{5}$ | $265{ }^{5}$ | $51^{\text {/ }}$ |  |  |
| Total | 1905 | 3479 | 3097 | 1412 | 568 | 426 | 404 | 570 | 1008 | 1195 | 2614 | 3932 | 3531 | $210^{5}$ |  |  |
| Used in assessment |  |  |  |  | 289 | 297 | 154 | 266 | 725 | 601 | 2106 | 3594 | 3239 | 1350 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1089 |  |  |
|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| Faroe Islands |  | 1094 | 1840 | 5957 | 3607 | 1270 | 1005 | 471 | 231 | 81 | 111 | 393 | 115 | 40 | 40 | 18 |
| Norw ay | 49 | 51 | 25 | 72 | 18 | 37 | 10 | 7 | 1 | 4 | 1 |  | 0 |  |  | 0 |
| Greenland | - | - | - | - | - | - | - | - | - | - | 5 |  | 1 |  |  |  |
| UK (EW/NI) | $18^{\text {5 }}$ | $50^{\text {² }}$ | $42^{5}$ | $15^{5}$ | $15^{\text {³ }}$ | $24^{\frac{5}{3}}$ | $1^{5}$ |  |  |  |  |  |  |  |  |  |
| UK (Scotland) | $245{ }^{\text {5 }}$ | $288{ }^{\frac{5}{3}}$ | $218{ }^{5}$ | $254{ }^{5}$ | $244{ }^{\frac{5}{3}}$ | $1129{ }^{\text {3 }}$ | $278{ }^{5}$ | 53 | 32 | 38 | 54 |  |  |  | 45 |  |
| Total | 312 | 1483 | 2125 | 6298 | 3884 | 2460 | 1294 | 531 | 264 | 123 | 171 | 393 | 116 | 40 | 85 |  |
| Correction of Faroese catches in Vb2 |  | -65 | -109 | -353 | -214 | -75 | -60 | -28 | -14 | -5 | -7 | -23 | -7 | -2 | -2 | -1 |
| Used in assessment | 1194 | 1080 | 1756 | 5676 | 3411 | 1232 | 955 | 450 | 218 | 80 | 105 | 370 | 108 | 38 | 38 | 17 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Preliminary |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{2}$ Included in Vb1. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{3}$ Reported as Vb. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 3.2. Faroe Bank (subdivision Vb2) cod. Surplus production model output using summer index.
Faroe Bank Cod RV Page 1

28 Apr 2016 at 13:32.23
ASPIC -- A Surplus-Production Model Including Covariates (Ver. 3.82)

## FIT Mode

ASPIC User's Manual is available gratis 101 Pivers Island Road; Beaufort, North Carolina 28516 USA
from the author.
Ref: Prager, M. H. 1994. A suite of extensions to a nonequilibrium surplus-production model. Fishery Bulletin 92: 374-389.

CONTROL PARAMETERS USED (FROM INPUT FILE)

| ----------------------------------------------------------------------------------------------------- |  |  |  |
| :--- | :---: | :---: | :---: | :---: |
| Number of years analyzed: |  | Number of bootstrap trials: | 0 |
| Number of dataseries: | 1 | Lower bound on MSY: | $5.000 \mathrm{E}+02$ |
| Objective function computed: | in effort | Upper bound on MSY: | $1.000 \mathrm{E}+09$ |
| Relative conv. criterion (simplex): | $1.000 \mathrm{E}-08$ | Lower bound on r: | $7.000 \mathrm{E}-02$ |
| Relative conv. criterion (restart): | $3.000 \mathrm{E}-08$ | Upper bound on r: | $2.500 \mathrm{E}+00$ |
| Relative conv. criterion (effort): | $1.000 \mathrm{E}-04$ | Random number seed: | 2010417 |
| Maximum F allowed in fitting: | 8.000 | Monte Carlo search mode, trials: | 1000 |

PROGRAM STATUS INFORMATION (NON-BOOTSTRAPPED ANALYSIS)
code 0
Normal convergence.

GOODNESS-OF-FIT AND WEIGHTING FOR NON-BOOTSTRAPPED ANALYSIS

| Weighted <br> Loss component number and title | Weighted SSE N | Current <br> MSE | Suggested weight | R-squared weight in | cpue |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |
| Loss(-1) SSE in yield 0.0 | $0.000 \mathrm{E}+00$ |  |  |  |  |
| Loss( 0) Penalty for B1R > 2 | $0.000 \mathrm{E}+00 \quad 1$ | N/A | $1.000 \mathrm{E}-01$ | N/A |  |
| Loss( 1) Survey cpue Spring | $1.545 \mathrm{E}+0133$ | $4.983 \mathrm{E}-01$ | $11.000 \mathrm{E}+00$ | $1.000 \mathrm{E}+00$ | 0.443 |
| TOTAL OBJECTIVE FUNCTION: | 1.5448103 | E+01 |  |  |  |
| Number of restarts required for convergence | gence: 10 |  |  |  |  |
| Est. B-ratio coverage index (0 worst, 2 best) | best): 0.7964 | < The | ese two measur | ures are defined | Prager |
| Est. B-ratio nearness index ( 0 worst, 1 best): | best): 0.8438 | < et | et al. (1996), Tr | Trans. A.F.S. 12 | 729 |

## MODEL PARAMETER ESTIMATES (NON-BOOTSTRAPPED)



## MANAGEMENT PARAMETER ESTIMATES (NON-BOOTSTRAPPED)


........ Fishing effort at MSY in units of each fishery:
fmsy ( 1) Survey cpue Spring $\quad 8.004 \mathrm{E}+00 \quad \mathrm{r} / 2 \mathrm{q}(1) \quad \mathrm{f}(0.1)=7.203 \mathrm{E}+00$

Page 2

ESTIMATED POPULATION TRAJECTORY (NON-BOOTSTRAPPED)


Faroe Bank Cod RV
Page 3

RESULTS FOR DATASERIES \# 1 (NON-BOOTSTRAPPED)
Survey cpue Spring


[^0]Faroe Bank Cod RV
Page 4
UNWEIGHTED LOG RESIDUAL PLOT FOR DATASERIES \# 1


Observed (O) and Estimated (*) cpue for Dataseries \# 1 -- Survey cpue Spring


Table 3.3. Faroe Bank (subdivision Vb2) cod. Surplus production model output using spring index.

Faroe Bank Cod RV | Page 1 |
| :--- |
| 28 Apr 2016 at 13:24.32 |

ASPIC -- A Surplus-Production Model Including Covariates (Ver. 3.82)
FIT Mode

Author: Michael H. Prager; NOAA/NMFS/S.E. Fisheries Science Center
101 Pivers Island Road; Beaufort, North Carolina 28516 USA
ASPIC User's Manual
is available gratis
from the author.
Ref: Prager, M. H. 1994. A suite of extensions to a nonequilibrium surplus-production model. Fishery Bulletin 92: 374-389.

CONTROL PARAMETERS USED (FROM INPUT FILE)

| Number of years analyzed: | 51 | Number of bootstrap trials: | 0 |
| :---: | :---: | :---: | :---: |
| Number of dataseries: | 1 | Lower bound on MSY: | $5.000 \mathrm{E}+02$ |
| Objective function computed: | in effort | Upper bound on MSY: | $1.000 \mathrm{E}+09$ |
| Relative conv. criterion (simplex): | $1.000 \mathrm{E}-08$ | Lower bound on r : | $7.000 \mathrm{E}-02$ |
| Relative conv. criterion (restart): | $3.000 \mathrm{E}-08$ | Upper bound on r : | $2.500 \mathrm{E}+00$ |
| Relative conv. criterion (effort): | $1.000 \mathrm{E}-04$ | Random number seed: | 2010417 |
| Maximum F allowed in fitting: | 8.000 | Monte Carlo search mode, trials: | 110000 |

PROGRAM STATUS INFORMATION (NON-BOOTSTRAPPED ANALYSIS)

Normal convergence.

GOODNESS-OF-FIT AND WEIGHTING FOR NON-BOOTSTRAPPED ANALYSIS

| Weighted | Weighted | Current | Suggested | R -squared |
| :---: | :---: | :---: | :---: | :---: |
| Loss component number and title | SSE N | MSE | weight | weight |

Loss(-1) SSE in yield
Loss( 0) Penalty for B1R > 2
Loss( 1) Survey cpue Spring
TOTAL OBJECTIVE FUNCTION:
$0.000 \mathrm{E}+00$
$0.000 \mathrm{E}+00 \quad 1 \quad$ N/A $\quad 1.000 \mathrm{E}-01 \quad$ N/A
$2.279 \mathrm{E}+01 \quad 30 \quad 8.140 \mathrm{E}-01 \quad 1.000 \mathrm{E}+00 \quad 1.000 \mathrm{E}+00 \quad 0.102$ $2.27906013 \mathrm{E}+01$

Number of restarts required for convergence: 33
Est. B-ratio coverage index ( 0 worst, 2 best): 0.3202
< These two measures are defined in Prager
Est. B-ratio nearness index (0 worst, 1 best): $0.3595<$ et al. (1996), Trans. A.F.S. 125:729


MANAGEMENT PARAMETER ESTIMATES (NON-BOOTSTRAPPED)

| Parameter |  | Estimate Formula | Related quantity |
| :---: | :---: | :---: | :---: |
| MSY | Maximum sustainable yield | - $4.789 \mathrm{E}+03$ | $\mathrm{Kr} / 4$ |
| K | Maximum stock biomass | $5.756 \mathrm{E}+04$ |  |
| Bmsy | Stock biomass at MSY | $2.878 \mathrm{E}+04$ | K/2 |
| Fmsy | Fishing mortality at MSY | $1.664 \mathrm{E}-01$ | r/2 |
| $\mathrm{F}(0.1)$ | Management benchmark | $1.498 \mathrm{E}-01$ | 0.9*Fmsy |
| $\mathrm{Y}(0.1)$ | Equilibrium yield at $\mathrm{F}(0.1)$ | $4.741 \mathrm{E}+03$ | 0.99*MSY |
| B-ratio | Ratio of B(2016) to Bmsy | $2.953 \mathrm{E}-01$ |  |
| F-ratio | Ratio of F(2015) to Fmsy | $1.382 \mathrm{E}-02$ |  |
| F01-mu | lt Ratio of $\mathrm{F}(0.1)$ to $\mathrm{F}(2015)$ | $6.513 \mathrm{E}+01$ |  |
| Y-ratio | Proportion of MSY avail in 20 | $2016 \quad 5.034 \mathrm{E}-01$ | $2 * \mathrm{Br}^{-\mathrm{Br}^{\wedge} 2} \mathrm{Ye}(20$ |

........ Fishing effort at MSY in units of each fishery:
fmsy (1) Survey cpue Spring $\quad 5.973 \mathrm{E}+00 \quad \mathrm{r} / 2 \mathrm{q}(1) \quad \mathrm{f}(0.1)=5.376 \mathrm{E}+00$

Page 2

## ESTIMATED POPULATION TRAJECTORY (NON-BOOTSTRAPPED)



| 1965 | 0.273 | $8.530 \mathrm{E}+03$ | $8.574 \mathrm{E}+03$ | $2.341 \mathrm{E}+03$ | $2.341 \mathrm{E}+03$ | $2.428 \mathrm{E}+03$ | $1.641 \mathrm{E}+00$ | $2.964 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 1966 | 0.214 | $8.618 \mathrm{E}+03$ | $8.916 \mathrm{E}+03$ | $1.909 \mathrm{E}+03$ | $1.909 \mathrm{E}+03$ | $2.507 \mathrm{E}+03$ | $1.287 \mathrm{E}+00$ | $2.994 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 1967 | 0.160 | $9.216 \mathrm{E}+03$ | $9.776 \mathrm{E}+03$ | $1.569 \mathrm{E}+03$ | $1.569 \mathrm{E}+03$ | $2.700 \mathrm{E}+03$ | $9.646 \mathrm{E}-01$ | $3.202 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 1968 | 0.397 | $1.035 \mathrm{E}+04$ | $9.741 \mathrm{E}+03$ | $3.871 \mathrm{E}+03$ | $3.871 \mathrm{E}+03$ | $2.692 \mathrm{E}+03$ | $2.388 \mathrm{E}+00$ | $3.595 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 5 | 1969 | 0.266 | $9.169 \mathrm{E}+03$ | $9.230 \mathrm{E}+03$ | $2.457 \mathrm{E}+03$ | $2.457 \mathrm{E}+03$ | $2.579 \mathrm{E}+03$ | $1.600 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 3.186 \mathrm{E}-01$


| 1970 | 0.332 | $9.291 \mathrm{E}+03$ | $9.055 \mathrm{E}+03$ | $3.002 \mathrm{E}+03$ | $3.002 \mathrm{E}+03$ | $2.539 \mathrm{E}+03$ | $1.992 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 3.228 \mathrm{E}-01$


| 1971 | 0.230 | $8.828 \mathrm{E}+03$ | $9.058 \mathrm{E}+03$ | $2.079 \mathrm{E}+03$ | $2.079 \mathrm{E}+03$ | $2.540 \mathrm{E}+03$ | $1.379 \mathrm{E}+00$ | $3.067 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 1972 | 0.228 | $9.289 \mathrm{E}+03$ | $9.528 \mathrm{E}+03$ | $2.168 \mathrm{E}+03$ | $2.168 \mathrm{E}+03$ | $2.646 \mathrm{E}+03$ | $1.368 \mathrm{E}+00$ | $3.227 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 9 | 1973 | 0.614 | $9.767 \mathrm{E}+03$ | $8.312 \mathrm{E}+03$ | $5.101 \mathrm{E}+03$ | $5.101 \mathrm{E}+03$ | $2.363 \mathrm{E}+03$ | $3.688 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $3.393 \mathrm{E}-01$ |  |  |  |  |  |  |  |  |


| 10 | 1974 | 0.295 | $7.029 \mathrm{E}+03$ | $7.020 \mathrm{E}+03$ | $2.068 \mathrm{E}+03$ | $2.068 \mathrm{E}+03$ | $2.051 \mathrm{E}+03$ | $1.770 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.442 \mathrm{E}-01$


| 11 | 1975 | 0.290 | $7.012 \mathrm{E}+03$ | $7.019 \mathrm{E}+03$ | $2.036 \mathrm{E}+03$ | $2.036 \mathrm{E}+03$ | $2.051 \mathrm{E}+03$ | $1.743 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | $2.436 \mathrm{E}-01$


| 12 | 1976 | 0.327 | $7.027 \mathrm{E}+03$ | $6.908 \mathrm{E}+03$ | $2.258 \mathrm{E}+03$ | $2.258 \mathrm{E}+03$ | $2.023 \mathrm{E}+03$ | $1.964 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.442 \mathrm{E}-01$


| 13 | 1977 | 0.130 | $6.792 \mathrm{E}+03$ | $7.370 \mathrm{E}+03$ | $9.590 \mathrm{E}+02$ | $9.590 \mathrm{E}+02$ | $2.138 \mathrm{E}+03$ | $7.821 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.360 \mathrm{E}-01$


| 4 | 1978 | 0.655 | $7.971 \mathrm{E}+03$ | $6.683 \mathrm{E}+03$ | $4.379 \mathrm{E}+03$ | $4.379 \mathrm{E}+03$ | $1.963 \mathrm{E}+03$ | $3.938 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.769 \mathrm{E}-01$


| 5 | 1979 | 0.227 | $5.555 E+03$ | $5.763 \mathrm{E}+03$ | $1.306 \mathrm{E}+03$ | $1.306 \mathrm{E}+03$ | $1.726 \mathrm{E}+03$ | $1.362 \mathrm{E}+00$ | $1.930 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 16 | 1980 | 0.191 | $5.975 \mathrm{E}+03$ | $6.303 \mathrm{E}+03$ | $1.203 \mathrm{E}+03$ | $1.203 \mathrm{E}+03$ | $1.868 \mathrm{E}+03$ | $1.147 \mathrm{E}+00$ | $2.076 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 17 | 1981 | 0.174 | $6.639 \mathrm{E}+03$ | $7.048 \mathrm{E}+03$ | $1.229 \mathrm{E}+03$ | $1.229 \mathrm{E}+03$ | $2.058 \mathrm{E}+03$ | $1.048 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.307 \mathrm{E}-01$


| 18 | 1982 | 0.293 | $7.468 \mathrm{E}+03$ | $7.456 \mathrm{E}+03$ | $2.184 \mathrm{E}+03$ | $2.184 \mathrm{E}+03$ | $2.160 \mathrm{E}+03$ | $1.760 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.595 \mathrm{E}-01$


| 19 | 1983 | 0.310 | $7.444 \mathrm{E}+03$ | $7.370 \mathrm{E}+03$ | $2.284 \mathrm{E}+03$ | $2.284 \mathrm{E}+03$ | $2.139 \mathrm{E}+03$ | $1.862 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.586 \mathrm{E}-01$



| 21 | 1985 | 0.432 | $7.221 \mathrm{E}+03$ | $6.740 \mathrm{E}+03$ | $2.913 \mathrm{E}+03$ | $2.913 \mathrm{E}+03$ | $1.980 \mathrm{E}+03$ | $2.597 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | $2.509 \mathrm{E}-01$


| 22 | 1986 | 0.291 | $6.288 \mathrm{E}+03$ | $6.304 \mathrm{E}+03$ | $1.836 \mathrm{E}+03$ | $1.836 \mathrm{E}+03$ | $1.868 \mathrm{E}+03$ | $1.750 \mathrm{E}+00$ | $2.185 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 23 | 1987 | 0.635 | $6.320 \mathrm{E}+03$ | $5.370 \mathrm{E}+03$ | $3.409 \mathrm{E}+03$ | $3.409 \mathrm{E}+03$ | $1.619 \mathrm{E}+03$ | $3.815 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.196 \mathrm{E}-01$


| 24 | 1988 | 0.846 | $4.529 \mathrm{E}+03$ | $3.508 \mathrm{E}+03$ | $2.966 \mathrm{E}+03$ | $2.966 \mathrm{E}+03$ | $1.094 \mathrm{E}+03$ | $5.082 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | $1.574 \mathrm{E}-01$


| 25 | 1989 | 0.530 | $2.658 \mathrm{E}+03$ | $2.395 \mathrm{E}+03$ | $1.270 \mathrm{E}+03$ | $1.270 \mathrm{E}+03$ | $7.638 \mathrm{E}+02$ | $3.186 \mathrm{E}+00$ | $9.235 \mathrm{E}-02$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 26 | 1990 | 0.121 | $2.152 \mathrm{E}+03$ | $2.380 \mathrm{E}+03$ | $2.890 \mathrm{E}+02$ | $2.890 \mathrm{E}+02$ | $7.590 \mathrm{E}+02$ | $7.299 \mathrm{E}-01$ | $7.476 \mathrm{E}-02$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 27 | 1991 | 0.102 | $2.622 \mathrm{E}+03$ | $2.925 \mathrm{E}+03$ | $2.970 \mathrm{E}+02$ | $2.970 \mathrm{E}+02$ | $9.237 \mathrm{E}+02$ | $6.103 \mathrm{E}-01$ | $9.110 \mathrm{E}-02$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

$28 \quad 1992 \quad 0.041 \quad 3.249 \mathrm{E}+03 \quad 3.731 \mathrm{E}+03 \quad 1.540 \mathrm{E}+02 \quad 1.540 \mathrm{E}+02 \quad 1.161 \mathrm{E}+03 \quad 2.481 \mathrm{E}-01 \quad 1.129 \mathrm{E}-01$
$29 \quad 1993 \quad 0.055 \quad 4.255 \mathrm{E}+03 \quad 4.837 \mathrm{E}+03 \quad 2.660 \mathrm{E}+02 \quad 2.660 \mathrm{E}+02 \quad 1.474 \mathrm{E}+03 \quad 3.305 \mathrm{E}-01 \quad 1.478 \mathrm{E}-01$

| 30 | 1994 | 0.121 | $5.463 \mathrm{E}+03$ | $5.979 \mathrm{E}+03$ | $7.250 \mathrm{E}+02$ | $7.250 \mathrm{E}+02$ | $1.782 \mathrm{E}+03$ | $7.288 \mathrm{E}-01$ | $1.898 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 31 | 1995 | 0.083 | $6.520 \mathrm{E}+03$ | $7.253 \mathrm{E}+03$ | $6.010 \mathrm{E}+02$ | $6.010 \mathrm{E}+02$ | $2.108 \mathrm{E}+03$ | $4.980 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.265 \mathrm{E}-01$


| 32 | 1996 | 0.259 | $8.028 \mathrm{E}+03$ | $8.138 \mathrm{E}+03$ | $2.106 \mathrm{E}+03$ | $2.106 \mathrm{E}+03$ | $2.325 \mathrm{E}+03$ | $1.555 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.789 \mathrm{E}-01$


| 33 | 1997 | 0.479 | $8.247 \mathrm{E}+03$ | $7.508 \mathrm{E}+03$ | $3.594 \mathrm{E}+03$ | $3.594 \mathrm{E}+03$ | $2.172 \mathrm{E}+03$ | $2.877 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.865 \mathrm{E}-01$


| 34 | 1998 | 0.533 | $6.825 \mathrm{E}+03$ | $6.077 \mathrm{E}+03$ | $3.239 \mathrm{E}+03$ | $3.239 \mathrm{E}+03$ | $1.808 \mathrm{E}+03$ | $3.203 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.371 \mathrm{E}-01$


| 35 | 1999 | 0.174 | $5.393 \mathrm{E}+03$ | $5.748 \mathrm{E}+03$ | $1.001 \mathrm{E}+03$ | $1.001 \mathrm{E}+03$ | $1.722 \mathrm{E}+03$ | $1.047 \mathrm{E}+00$ | $1.874 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 36 | 2000 | 0.185 | $6.114 \mathrm{E}+03$ | $6.468 \mathrm{E}+03$ | $1.194 \mathrm{E}+03$ | $1.194 \mathrm{E}+03$ | $1.910 \mathrm{E}+03$ | $1.110 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |$\quad 2.124 \mathrm{E}-01$


| 37 | 2001 | 0.147 | $6.830 \mathrm{E}+03$ | $7.347 \mathrm{E}+03$ | $1.080 \mathrm{E}+03$ | $1.080 \mathrm{E}+03$ | $2.132 \mathrm{E}+03$ | $8.834 \mathrm{E}-01$ | $2.373 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 38 | 2002 | 0.215 | $7.882 \mathrm{E}+03$ | $8.170 \mathrm{E}+03$ | $1.756 \mathrm{E}+03$ | $1.756 \mathrm{E}+03$ | $2.333 \mathrm{E}+03$ | $1.292 \mathrm{E}+00$ | $2.739 \mathrm{E}-01$ |
| 39 | 2003 | 0.892 | $8.459 \mathrm{E}+03$ | $6.360 \mathrm{E}+03$ | $5.676 \mathrm{E}+03$ | $5.676 \mathrm{E}+03$ | $1.876 \mathrm{E}+03$ | $5.364 \mathrm{E}+00$ | $2.939 \mathrm{E}-01$ |
| 40 | 2004 | 1.024 | $4.659 \mathrm{E}+03$ | $3.331 \mathrm{E}+03$ | $3.411 \mathrm{E}+03$ | $3.411 \mathrm{E}+03$ | $1.042 \mathrm{E}+03$ | $6.153 \mathrm{E}+00$ | $1.619 \mathrm{E}-01$ |
| 41 | 2005 | 0.624 | $2.290 \mathrm{E}+03$ | $1.975 \mathrm{E}+03$ | $1.232 \mathrm{E}+03$ | $1.232 \mathrm{E}+03$ | $6.347 \mathrm{E}+02$ | $3.748 \mathrm{E}+00$ | $7.956 \mathrm{E}-02$ |
| 42 | 2006 | 0.666 | $1.692 \mathrm{E}+03$ | $1.433 \mathrm{E}+03$ | $9.550 \mathrm{E}+02$ | $9.550 \mathrm{E}+02$ | $4.649 \mathrm{E}+02$ | $4.005 \mathrm{E}+00$ | $5.880 \mathrm{E}-02$ |
| 43 | 2007 | 0.386 | $1.202 \mathrm{E}+03$ | $1.167 \mathrm{E}+03$ | $4.500 \mathrm{E}+02$ | $4.500 \mathrm{E}+02$ | $3.805 \mathrm{E}+02$ | $2.317 \mathrm{E}+00$ | $4.177 \mathrm{E}-02$ |
| 44 | 2008 | 0.179 | $1.133 \mathrm{E}+03$ | $1.221 \mathrm{E}+03$ | $2.180 \mathrm{E}+02$ | $2.180 \mathrm{E}+02$ | $3.975 \mathrm{E}+02$ | $1.073 \mathrm{E}+00$ | $3.936 \mathrm{E}-02$ |
| 45 | 2009 | 0.053 | $1.312 \mathrm{E}+03$ | $1.508 \mathrm{E}+03$ | $8.000 \mathrm{E}+01$ | $8.000 \mathrm{E}+01$ | $4.885 \mathrm{E}+02$ | $3.189 \mathrm{E}-01$ | $4.560 \mathrm{E}-02$ |
| 46 | 2010 | 0.053 | $1.721 \mathrm{E}+03$ | $1.974 \mathrm{E}+03$ | $1.050 \mathrm{E}+02$ | $1.050 \mathrm{E}+02$ | $6.343 \mathrm{E}+02$ | $3.196 \mathrm{E}-01$ | $5.979 \mathrm{E}-02$ |
| 47 | 2011 | 0.151 | $2.250 \mathrm{E}+03$ | $2.450 \mathrm{E}+03$ | $3.700 \mathrm{E}+02$ | $3.700 \mathrm{E}+02$ | $7.806 \mathrm{E}+02$ | $9.075 \mathrm{E}-01$ | $7.818 \mathrm{E}-02$ |
| 48 | 2012 | 0.035 | $2.661 \mathrm{E}+03$ | $3.072 \mathrm{E}+03$ | $1.080 \mathrm{E}+02$ | $1.080 \mathrm{E}+02$ | $9.673 \mathrm{E}+02$ | $2.113 \mathrm{E}-01$ | $9.245 \mathrm{E}-02$ |
| 49 | 2013 | 0.009 | $3.520 \mathrm{E}+03$ | $4.107 \mathrm{E}+03$ | $3.800 \mathrm{E}+01$ | $3.800 \mathrm{E}+01$ | $1.269 \mathrm{E}+03$ | $5.561 \mathrm{E}-02$ | $1.223 \mathrm{E}-01$ |
| 50 | 2014 | 0.007 | $4.751 \mathrm{E}+03$ | $5.527 \mathrm{E}+03$ | $3.800 \mathrm{E}+01$ | $3.800 \mathrm{E}+01$ | $1.661 \mathrm{E}+03$ | $4.132 \mathrm{E}-02$ | $1.651 \mathrm{E}-01$ |
| 51 | 2015 | 0.002 | $6.374 \mathrm{E}+03$ | $7.393 \mathrm{E}+03$ | $1.700 \mathrm{E}+01$ | $1.700 \mathrm{E}+01$ | $2.142 \mathrm{E}+03$ | $1.382 \mathrm{E}-02$ | $2.215 \mathrm{E}-01$ |
| 52 | 2016 | $8.499 \mathrm{E}+03$ |  |  |  |  | $2.953 \mathrm{E}-01$ |  |  |

Data type CC: cpue-catch series Series weight: 1.000


| 37 | 2001 | $1.022 \mathrm{E}+03$ | $2.047 \mathrm{E}+02$ | 0.1470 | $1.080 \mathrm{E}+03$ | $1.080 \mathrm{E}+03$ | -1.60839 | $0.000 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 38 | 2002 | $4.439 \mathrm{E}+02$ | $2.276 \mathrm{E}+02$ | 0.2149 | $1.756 \mathrm{E}+03$ | $1.756 \mathrm{E}+03$ | -0.66803 | $0.000 \mathrm{E}+00$ |
| 39 | 2003 | $8.671 \mathrm{E}+02$ | $1.772 \mathrm{E}+02$ | 0.8925 | $5.676 \mathrm{E}+03$ | $5.676 \mathrm{E}+03$ | -1.58807 | $0.000 \mathrm{E}+00$ |
| 40 | 2004 | $*$ | $9.281 \mathrm{E}+01$ | 1.0239 | $3.411 \mathrm{E}+03$ | $3.411 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 41 | 2005 | $*$ | $5.503 \mathrm{E}+01$ | 0.6236 | $1.232 \mathrm{E}+03$ | $1.232 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 42 | 2006 | $6.051 \mathrm{E}+01$ | $3.992 \mathrm{E}+01$ | 0.6664 | $9.550 \mathrm{E}+02$ | $9.550 \mathrm{E}+02$ | -0.41589 | $0.000 \mathrm{E}+00$ |
| 43 | 2007 | $5.206 \mathrm{E}+01$ | $3.251 \mathrm{E}+01$ | 0.3855 | $4.500 \mathrm{E}+02$ | $4.500 \mathrm{E}+02$ | -0.47070 | $0.000 \mathrm{E}+00$ |
| 44 | 2008 | $6.402 \mathrm{E}+01$ | $3.400 \mathrm{E}+01$ | 0.1786 | $2.180 \mathrm{E}+02$ | $2.180 \mathrm{E}+02$ | -0.63283 | $0.000 \mathrm{E}+00$ |
| 45 | 2009 | $5.550 \mathrm{E}+01$ | $4.200 \mathrm{E}+01$ | 0.0531 | $8.000 \mathrm{E}+01$ | $8.000 \mathrm{E}+01$ | -0.27870 | $0.000 \mathrm{E}+00$ |
| 46 | 2010 | $5.808 \mathrm{E}+01$ | $5.500 \mathrm{E}+01$ | 0.0532 | $1.050 \mathrm{E}+02$ | $1.050 \mathrm{E}+02$ | -0.05455 | $0.000 \mathrm{E}+00$ |
| 47 | 2011 | $1.224 \mathrm{E}+02$ | $6.826 \mathrm{E}+01$ | 0.1510 | $3.700 \mathrm{E}+02$ | $3.700 \mathrm{E}+02$ | -0.58401 | $0.000 \mathrm{E}+00$ |
| 48 | 2012 | $4.454 \mathrm{E}+01$ | $8.557 \mathrm{E}+01$ | 0.0352 | $1.080 \mathrm{E}+02$ | $1.080 \mathrm{E}+02$ | 0.65294 | $0.000 \mathrm{E}+00$ |
| 49 | 2013 | $1.390 \mathrm{E}+02$ | $1.144 \mathrm{E}+02$ | 0.0093 | $3.800 \mathrm{E}+01$ | $3.800 \mathrm{E}+01$ | -0.19452 | $0.000 \mathrm{E}+00$ |
| 50 | 2014 | $2.092 \mathrm{E}+02$ | $1.540 \mathrm{E}+02$ | 0.0069 | $3.800 \mathrm{E}+01$ | $3.800 \mathrm{E}+01$ | -0.30657 | $0.000 \mathrm{E}+00$ |
| 51 | 2015 | $3.719 \mathrm{E}+01$ | $2.060 \mathrm{E}+02$ | 0.0023 | $1.700 \mathrm{E}+01$ | $1.700 \mathrm{E}+01$ | 1.71162 | $0.000 \mathrm{E}+00$ |

* Asterisk indicates missing value(s).


## UNWEIGHTED LOG RESIDUAL PLOT FOR DATASERIES \# 1




```
Faroe Bank Cod RV
Observed (O) and Estimated (*) cpue for Dataseries \# 1 -- Survey cpue Spring
1200. -:
O
1000. -:
\(:\)
.
O
800. -:
600. -:
O O
:
\(\stackrel{:}{.}\)
```



```
400. -:
```



```
0. -:
```



```
1959. 1965. 1971. 1977. 1983. 1989. 1995. 2001. 2007. 2013. 2019.
                                    Time Plot of Estimated F-Ratio and B-Ratio
7.2 -:
    :
    :
    : F
6.0 -:
    :
    : F
    F
4.8-:
    :
    : F
            F F F
3.6-: F
    : F F
                            F
                                    F
2.4 -: F F
        F F FF
    F F FF F F
        FF F
1.2-: F F F F
    : -- --- --2 --- -- --- -- - 2- --- --- --- -- --- - 2- 2--- -- - 2- - 2- --- --- --
                                    FF F
                                    F
        B BB B BB B BB B BB B BB B BB B BB B BB F 2B B BB B BB B BB B FF F B BB B
        0.0 -: B BB B BB B BB B BB F FF
    1959. 1965. 1971. 1977. 1983. 1989. 1995. 2001. 2007. 2013. 2019.
```



Figure 3.1. Faroe Bank (subdivision Vb2) cod. Reported landings 1965-2015. Since 1992 only catches from Faroese and Norwegian vessels are considered to be taken on Faroe Bank. Lower plot: fishing days (fishing year) 1997-2016 for longline gear type in the Faroe Bank.


Figure 3.2. Faroe Bank (subdivision Vb 2 ) cod. Catch per unit of effort in spring groundfish survey (1983-2016)(red line) and summer survey (1996-2015)(black line). Vertical bars and shaded areas show the standard error in the estimation of indices.


Figure 3.3. Faroe Bank (subdivision Vb2) cod. Length distributions in summer survey (1996-2015)


Figure 3.4. Faroe Bank (subdivision Vb2) cod. Age-disaggregated indices in summer survey (ages 1-10)(1996-2015)


Figure 3.5. Faroe Bank (subdivision Vb2) cod. Length distributions in spring survey (1994-2016). No surveys were conducted in 1996, 2004 and 2005.


Figure 3.6. Faroe Bank (subdivision Vb2) cod. Age-disaggregated indices in spring survey (ages 110) (1994-2015). No surveys were conducted in 1996, 2004 and 2005. Data for 2016 were not available due to lack of age readings.

Recruitment yearclasses of Faroe Bank cod
(correlation from 1995 to 2014 equals 0.86 )


Figure 3.7. Faroe Bank (subdivision Vb 2 ) cod. Correlation between recruitment year classes in both survey indices.


Figure 3.8. Results from the surplus production model using summer index. Observed (points) and expected catch rates ( $\mathrm{kg} / \mathrm{hour}$ ) (top panel). Estimated fishing mortality (black line) and exploitation ratios (ratio of spring index to landings)(green line) (ratio of summer index to landings)(red line)(middle panel). Model residuals in log scale (bottom panel)


Figure 3.9. Results from the surplus production model using spring index. Observed (points) and expected catch rates (kg/hour) (top panel). Estimated fishing mortality (black line) and exploitation ratios (ratio of spring index to landings)(green line) (ratio of summer index to landings)(red line)(middle panel). Model residuals in log scale (bottom panel)


#### Abstract

Summary The input data consisted of the catch-at-age matrix (ages $2-10+$ years) for the period 1959-2015 and two age-disaggregated abundance indices obtained from the two Faroese groundfish surveys: spring survey 1994-2016 (shifted back to the previous year) and summer survey 1996-2015. The maturities were obtained from spring survey 19832016.

The assessment settings were the same as in the 2015 assessment. An XSA was run and tuned with the two survey indices. The fishing mortality in 2015 (average of ages 3-7 years) was estimated at 0.46 , which was higher than the FMSY of 0.32 . The total stock size (age $2+$ ) at the beginning of 2015 was estimated at 28500 tonnes and the spawningstock biomass at 19700 tonnes, which was slightly below the limit biomass of 21000 tonnes.

The short-term prediction until year 2018 showed a slightly decreasing total-stock biomass to 25000 tonnes and a spawning-stock biomass to 19000 tonnes.

It is advised to reduce the fishing mortality substantially to rebuild the stock.

\subsection*{4.1 Stock description and management units}

Both genetic and tagging data suggest that there are three cod stocks present in Faroese waters: on the Faroe Bank (Division 5.b.2), on the Faroe Plateau (Division 5.b.1) and on the Faroe-Iceland Ridge. Cod on the Faroe-Iceland Ridge seem to belong to the cod stock at Iceland, and the WG in 2005 decided to exclude these catches from the catch-at-age calculations. The annex provides more information.


### 4.2 Scientific data

### 4.2.1 Trends in landings and fisheries

The landings were obtained from the Fisheries Ministry and Statistics Faroe Islands. The landings are presented in Table 4.2.1 and the working group estimates are presented in Table 4.2.2. The catches on the Faroe-Iceland Ridge, i.e. for the large single trawlers and the large longliners were not included in the catch-at-age calculations. This year the catch figures back to 1999 on the Faroe-Iceland Ridge were revised. They were extracted from the database on the Faroese Coastal Guard directly using their definition of the relevant area. In recent years the longliners have taken the majority of the cod catches (Table 4.2.3).

### 4.2.2 Catch-at-age

Landings-at-age for 2015 are provided for the Faroese fishery in Table 4.2.4. Faroese landings from most of the fleet categories were sampled (Table 4.2.5). The catch-at-age is shown in Table 4.2.6. Catch curves are shown in Figure 4.2.1.

### 4.2.3 Weight-at-age

Mean weight-at-age data are provided for the Faroese fishery in Table 4.2.7. These were calculated using the length/weight relationship based on individual length/weight measurements of samples from the landings. The sum-of-products-check for 2015 showed a discrepancy of $0 \%$. The weights have increased in recent years (Figure 4.2.2).

### 4.2.4 Maturity-at-age

The proportion of mature cod by age during the Faroese groundfish surveys carried out during the spawning period (March) is given in Table 4.2.8 and in Figure 4.2.3. Full maturity is generally reached at age 5 or 6 , but considerable changes have been observed in the proportion mature for younger ages between years.

### 4.2.5 Catch, effort and research vessel data

Fisheries independent cpue series
Spring groundfish surveys in Faroese waters with the research vessel Magnus Heinason is used as a tuning series. The catch curves showed a normal pattern (Figure 4.2.4), i.e. a decreasing trend after age 5 . The stratified mean catch of cod per unit effort (Figure 4.2.5) has been low in the recent years.
The other tuning series used is summer Groundfish Survey. The stratified mean catch of cod per unit effort has been low in recent years (Figure 4.2.5). The catch curves (Figure 4.2.6) show that the fish are fully recruited to the survey gear at an age of 4 or 5 years. The YC2009 was present in nearly constant numbers in summer groundfish survey. Both tuning series are presented in Table 4.2 .9 and they show that there are few small cod in the stock.

Commercial cpue series
Three commercial cpue series (longliners and pairtrawlers) are also presented (Tables 4.2.10, 4.2.11, and 4.2.12 as well as Figure 4.2.7), although they are not used as tuning series. All these series show that the incoming year classes are small. Note that the small boats (0-25 GRT) operating with longlines and jigging reels close to land have had a relatively higher cpue in recent years compared with the other cpue series and the two tuning series (Figure 4.2.8 and Figure 4.2.9), although the larger longliners also have had a high catchability in recent years. When that happens, the recruitment of 2year old cod tends to be low.

### 4.3 Information from the fishing industry

The sampling of the catches is included in the 'scientific data'. The fishing industry has since 1996 gathered data on the size composition of the landings but this information has not been used in this assessment.

### 4.4 Methods

This is an update assessment using XSA and the procedure is described in stock annex and the results of the assessment is mostly data-driven implying that there may be little difference in the assessment results by using another method.

### 4.5 Reference points

The reference points are dealt with in the general section of Faroese stocks. The PA reference points for Faroe Plateau cod are the following: $\mathrm{B}_{\mathrm{pa}}=40 \mathrm{kt}$, Blim $=21 \mathrm{kt}, \mathrm{F}_{\mathrm{pa}}=$ 0.35 and $\mathrm{F}_{\mathrm{lim}}=0.68$.

The reference points based on the yield-per-recruit curve are the following: $\mathrm{F}_{\max }=0.25$, $\mathrm{F}_{0.1}=0.12, \mathrm{~F} 35 \% \mathrm{SPR}=0.18, \mathrm{~F}_{\text {med }}=0.37, \mathrm{~F}_{\text {low }}=0.10, \mathrm{~F}_{\text {high }}=0.91$.

The group adopted in 2011 following preliminary MSY reference points: $\mathrm{F}_{\text {msy }}=0.32$, see section 4.8. The $\mathrm{B}_{\text {trigger }}$ was set at $\mathrm{B}_{\mathrm{pa}}=40 \mathrm{kt}$.

### 4.6 State of the stock - historical and compared to what is now

Since the current assessment is an update assessment, the same procedure is followed as last year: to use the two surveys for tuning. The commercial series showed a similar overall tendency as the surveys (Figure 4.2.7) but were not used in the tuning. The XSA-run (Table 4.6.1) showed that the fit between the model and the tuning series (logQ residuals, Figure 4.6.1) was rather poor for the young ages and there seemed to be both year class effects and year effects.

The results from the XSA-run shows that fishing mortality (F3-7) has increased in recent years (Table 4.6.2, Figure 4.6.2), and other measures of fishing mortality have done so as well (Table 4.6.4, Figure 4.6.3). The population numbers, total biomass and spawning-stock biomass have been low compared with other years in the series (Table 4.6.3, Table 4.6.4, Figure 4.6.2). The poor state of the stock since 2005 has been due to poor recruitment (not poor individual growth). Prior to that time, extremely weak year classes ( $<5$ million individuals) were only observed two times, whereas it has happened four times since 2005. In the past there has been a poor relationship between the size of the spawning stock and subsequent recruitment (Figure 4.6.4), but the increasing number of low data points in recent years have strengthened the stock-recruitment relationship. The spawning-stock biomass in the terminal year was below Blim and the fishing mortality above $\mathrm{F}_{\mathrm{MSY}}$ and $\mathrm{Fpa}_{\mathrm{pa}}$ (Figure 4.6.5).
To put the recent low biomass of Faroe Plateau cod into a wider context, and also to provide a basis to evaluate ecosystem properties and biomass reference points, the stock size (age 2+) of Faroe Plateau cod has been estimated back to 1710 (Table 4.6.5, Table 4.6.6, Figure 4.6.6). The first step, estimating the 1906-1958 period, was done in NWWG 2015 whereas the rest was done this year. Scaling Shetland and Faroese cpue series 1859-1914 gave biomass estimates from 1860-1905 (Working Document 24). In the same period there was information about occasional 'good' and 'bad' cod years. It happened that 'bad' years corresponded to approximately 80 thousand tons of cod and 'good' years to approximately 220 thousand tons of cod. Information about occasional 'good' and 'bad' years were available back to the seventeenth century, i.e. the biomass was roughly known for these years. From 1709 to 1856 the Royal Monopoly Trade (Den Kongelige Monopolhandel) recorded export of dried cod from the Faroes almost every year. Based on this export and an effect of year (taking into account an increasing tendency to export dried cod over the years) gave estimates of cod for the 1710-1859 period (Working Document 25, Model 1). An attempt (Model 2, and a factor between dried cod export and biomass 1841-1856) was also made to take account of a possible increased access to landing places during the 1841-1856 period, i.e. that the relationship between dried cod export and biomass had changed compared with the 17101840 period. The biomass estimates were quite different for the 1841-1859 period and more work could therefore be done to evaluate the importance of the potential increased access to landing sites and other technological improvements in the fishery. The results prior to 1906 should be interpreted with care, especially prior to 1860 . The results indicate that the poor state of the stock in recent years is unprecedented the last three centuries.

### 4.7 Short-term forecast

### 4.7.1 Input data

The input data for the short-term prediction are given in Table 4.7.1. Note the extremely weak YC2013 and YC2014, which were set to the face value from the XSA-run,
i.e. according to the Annex. Estimates of stock size (ages 3+) were taken directly from the XSA stock numbers. The exploitation pattern was estimated as the average fishing mortality for 2013-2015 and scaled to the 2015 value. The weights at age in the catches in 2016 were estimated from spring survey (ages 2 and $6-8$ years) whereas the other ages were estimated from the catch weights in January-February 2016. The weights in the catches in 2017 were set to the values in 2016 and the average of 2014-2016 was expected for 2018. The proportion mature in 2016 was set to the 2016 values from spring groundfish survey, and for 2017-2018 to the average values for 2014-2016.

### 4.7.2 Results

The landings in 2016 are expected to be 7500 tonnes (Table 4.7.2) (the landings from the Faroe-Icelandic ridge should be added to this figure in order to get the total Faroese landings within the 5.b. 1 area). The spawning-stock biomass is expected to be 22400 tonnes in 2016, 20200 tonnes in 2017 and eventually 18900 tonnes in 2017. Many year classes contribute to the SSB in 2017-18 (Figure 4.7.1).

### 4.8 Long-term forecast

The input to the traditional long-term forecast (yield-per-recruit) is presented in Table 4.8.1 and the result is presented in Table 4.8.2 and Figure 4.8.1.

Single species long-term forecasts for Faroe Plateau cod indicated Fmsy values lower than $\mathrm{F}_{\mathrm{pa}}$. An FLR procedure (MSE, Management strategy evaluations using FLR standard packages; a simulation of management and stock response over a 20 yr period) for Faroe Plateau cod indicates that $\mathrm{F}_{\mathrm{mSY}}$ is 0.32 . This value ( 0.32 ) was adopted by the NWWG 2011 as a preliminary Fmsy.

### 4.9 Uncertainties in assessment and forecast

Since there is no incentive to discard fish or misreport catches under the effort management system, the catch figures are considered adequate, as well as the catch-at-age, although the number of otoliths should have been higher.

The retrospective pattern indicates less uncertainty in the assessment than seen some years ago (Figure 4.9.1).

Steingrund et al. (2010) found that the recruitment of Faroe Plateau cod (age 2) could be rather precisely estimated as there is a relationship between cod biomass (age 3+) and the amount of cannibalistic cod in nearshore waters in June-October the previous year. This approach showed that the recent year classes were extremely weak (Figure 4.9.2).

### 4.10 Comparison with previous assessment and forecast

The assessment settings were according to the Stock Annex. The 2016 assessment was much in line with the 2015 assessment and forecast (Figure 4.10.1).

### 4.11 Management plans and evaluations

There is no explicit management plan for this stock. A management system based on number of fishing days, closed areas and other technical measures was introduced in 1996 with the purpose to ensuring sustainable demersal fisheries in subarea 5.b. This was before ICES introduced PA and MSY reference values and at the time it was believed that the purpose was achieved, if the total allowable number of fishing days was
set such that on average $33 \%$ of the cod exploitable stock in numbers would be harvested annually. This translates into an average F of 0.45 , above the $\mathrm{F}_{\mathrm{pa}}$ of 0.35 . ICES considers this to be inconsistent with the PA and MSY approaches. Some work has been done in the Faroes to move away from the Ftarget of 0.45 to be more consistent with the ICES advice. A committee set by Faroese authorities to evaluate the management system will deliver its recommendations in summer 2016. The recommendations along with political modifications are expected to be set in force in 2018.

### 4.12 Management considerations

The cod stock is assessed to be in a very poor state and is predicted to remain so for the next two years due to poor recruitment. Although the environmental conditions have been rather special since 2007 (lots of mackerel) and may partly be responsible for the poor state of the cod stock, it is certainly necessary to protect the cod stock as much as possible. The reason is not only that it may prevent a total collapse of the stock but also that the stock may recover faster in future. Hence, a reduction in fishing mortality is urgently needed.

### 4.13 Ecosystem considerations

Regarding the ecosystem effects on fishing, this issue is partly addressed in the ecological modelling work presented in the overview section for Faroese stocks.

### 4.14 Regulations and their effects

There seems to be a poor relationship between the number of fishing days and the fishing mortality because of large fluctuations in catchability. Area restrictions may help to reduce fishing mortality, but they cause practical problems for the fishing fleets (e.g. high concentrations of vessels in certain areas). Area restrictions may be best suited to protect certain fish species/sizes in certain areas, whereas the number of fishing days remains the only tool to reduce the overall fishing mortality, given the effort management system.

The area closure (for commercial longliners close to land) introduced in July 2011 and ending in August 2013 to protect young fish has not yet resulted in strong recruitment, since the 2008 year class is below average size, and the 2009-2011 year classes either poor or exceptionally poor.

### 4.15 Changes in fishing technology and fishing patterns

Fishing effort per fishing day may have increased gradually since the effort management system was introduced in 1996, although little direct quantitative information exists. There also seems to have been substantial increases in fishing power when new vessels are replacing old vessels.

The fishing pattern in recent years has changed compared with previous years. The large longliners seem to have exploited the deep areas (> 200 m ) to a larger extent (ling and tusk) because the catches in shallower waters of cod and haddock have been so poor - which was also observed at the beginning of the 1990s. They also have fished in other areas, e.g. in Greenland and on the Flemish Cap. This could reduce the fishing mortality on cod and haddock, but the small longliners and jiggers still exploit the shallow areas.

### 4.16 Changes in the environment

The primary production has been low for a number of years, albeit high in 2008-2010, but it is not believed that this has any relationship with a change in the environment. The temperature has been high in recent years (although it has been a little bit cooler in 2014-2015), which may have a negative effect on cod recruitment (Planque and Fredou, 1999).

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### 4.18 Tables

Table 4.2.1. Faroe Plateau cod (subdivision 5.b.1). Nominal catch ( $\mathbf{t}$ ) by countries, as officially reported to ICES.

| Denmark Faroe Islands France |  |  |  | Germany | Iceland | Norway | Greenland Portugal | UK (E/W/NI) | UK (Scotland) | United Kingdom | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 8 | 34492 | 4 | 8 |  | 83 | - | 0 | 0 | - | 34595 |
| 1987 | 30 | 21303 | 17 | 12 |  | 21 | - | 8 | 0 | - | 21391 |
| 1988 | 10 | 22272 | 17 | 5 |  | 163 | - | 0 | 0 | - | 22467 |
| 1989 | - | 20535 | - | 7 |  | 285 | - | 0 | 0 | - | 20827 |
| 1990 | - | 12232 | - | 24 |  | 124 | - | 0 | 0 | - | 12380 |
| 1991 | - | 8203 |  | 16 |  | 89 | - | 1 | 0 | - | 8309 |
| 1992 | - | 5938 |  | 12 |  | 39 | - | 74 | 0 | - | 6066 |
| 1993 | - | 5744 | 1 | + |  | 57 | - | 186 | 0 | - | 5988 |
| 1994 | - | 8724 | - | $2 \cdots$ |  | 36 | - | 56 | 0 | - | 8818 |
| 1995 | - | 19079 |  | 2 |  | 38 | - | 43 | 0 | - | 19164 |
| 1996 | - | 39406 | 1 | + |  | 507 | - | 126 | 0 | - | 40040 |
| 1997 | - | 33556 | - | + |  | 410 | - | $61^{\cdots}$ | 0 | - | 34027 |
| 1998 | - | 23308 |  | - |  | 405 | - | $27^{\cdots}$ | 0 | - | 23740 |
| 1999 | - | 19156 |  | 39 | - | 450 | - | 51 | 0 |  | 19696 |
| 2000 |  | 0 | 1 | 2 | - | 374 | - | 18 | 0 |  | 395 |
| 2001 |  | 29762 |  | 9 | - | 531 * | - | 50 | 0 |  | 30361 |
| 2002 |  | 40602 | 20 | 6 | 5 | 573 |  | 42 | 0 |  | 41248 |
| 2003 |  | 30259 | 14 | 7 | - | 447 | - | 15 | 0 |  | 30742 |
| 2004 |  | 17540 | 2 | $3 \cdots$ |  | 414 | 1 | 15 | 0 |  | 17975 |
| 2005 |  | 13556 | - |  |  | 201 |  | 24 | 0 |  | 13781 |
| 2006 |  | 11629 | 7 | $1{ }^{\cdots}$ |  | 49 | 5 | 0 | 0 |  | 11691 |
| 2007 |  | 9905 |  |  |  | 71 | 7 | 0 | 360 |  | 10344 |
| 2008 |  | 9394 | 1 |  |  | 40 |  | 0 | 383 |  | 9818 |
| 2009 |  | 10736 | 1 |  |  | 14 | 7 | 0 | 300 |  | 11058 |
| 2010 |  | 13878 | 1 |  |  | 10.338 |  | 0 | 312 |  | 14201 |
| 2011 |  | 11348 | - |  |  | 0 |  | 0 | 0 |  | 11348 |
| 2012 |  | 8437 | 0 |  | 28 | 0 |  | 0 | 0 |  | 8465 |
| 2013 |  | 5331 | 0 |  | 20 | 0 | 2 | 0 | 0 |  | 5333 |
| 2014 |  | 6655 |  |  |  | 6.414 |  | 0 | 226 |  | 6887 |
| 2015 * |  | 7812 |  |  |  | 33 | 14 | 0 | 382 |  | 8241 |

Table 4.2.2. Faroe Plateau cod (subdivision 5.b.1). Nominal catch (t) used in the assessment.


Table 4.2.3. Faroe Plateau cod (subdivision 5.b.1). The landings of Faroese fleets (in percent) of total catch ( $\mathbf{t}$ ). Note that the catches on the Faroe-Iceland ridge (mainly belonging to single trawl-ers > 1000 HP ) are included in this table, but excluded in the XSA-run.

| Year |  | Open <br> boats |  | $\begin{aligned} & \text { Longliners } \\ & <100 \text { GRT } \end{aligned}$ | Singletrawl $<400 \mathrm{HP}$ | $\begin{aligned} & \text { Gill } \\ & \text { net } \end{aligned}$ |  | Jiggers |  | $\begin{aligned} & \hline \text { Singletrawl } \\ & 400-1000 \mathrm{HP} \end{aligned}$ | Singletrawl $>1000 \mathrm{HP}$ | $\begin{aligned} & \hline \text { Pairtrawl } \\ & <1000 \mathrm{HP} \end{aligned}$ | Pairtrawl $>1000 \mathrm{HP}$ | $\begin{aligned} & \text { Longliners } \\ & >100 \text { GRT } \end{aligned}$ | Industrial trawlers | Others |  | Faroe catch <br> Round.weight |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1985 |  | 16.0 | 27.2 |  | . 7 | 0.6 |  | 4.3 | 7.9 | 11.2 | 12.3 | 5.6 | 7.5 |  | 0.2 | 0.6 | 39,422 |
|  | 1986 |  | 9.5 | 15.1 |  | . 1 | 1.3 |  | 2.9 | 6.2 | 8.5 | 29.6 | 14.9 | 5.1 |  | 0.4 | 1.3 | 34,492 |
|  | 1987 |  | 9.9 | 14.8 |  | . 2 | 0.5 |  | 2.9 | 6.7 | 8.0 | 26.0 | 14.5 | 9.9 |  | 0.5 | 0.1 | 21,303 |
|  | 1988 |  | 2.6 | 13.8 |  | . 9 | 2.6 |  | 7.5 | 7.4 | 6.8 | 25.3 | 15.6 | 12.7 |  | 0.6 | 0.2 | 22,272 |
|  | 1989 |  | 4.4 | 29.0 |  | . 7 | 3.2 |  | 9.3 | 5.7 | 5.5 | 10.5 | 8.3 | 17.7 |  | 0.7 | 0.0 | 20,535 |
|  | 1990 |  | 3.9 | 35.5 |  | . 8 | 1.4 |  | 8.2 | 3.7 | 4.3 | 7.1 | 10.5 | 19.6 |  | 0.6 | 0.2 | 12,232 |
|  | 1991 |  | 4.3 | 31.6 |  | 7.1 | 2.0 |  | 8.0 | 3.4 | 4.7 | 8.3 | 12.9 | 17.2 |  | 0.6 | 0.1 | 8,203 |
|  | 1992 |  | 2.6 | 26.0 |  | . 9 | 0.0 |  | 7.0 | 2.2 | 3.6 | 12.0 | 20.8 | 13.4 |  | 5.0 | 0.4 | 5,938 |
|  | 1993 |  | 2.2 | 16.0 | 15. |  | 0.0 |  | 9.0 | 4.1 | 3.6 | 14.2 | 21.7 | 12.6 |  | 0.8 | 0.4 | 5,744 |
|  | 1994 |  | 3.1 | 13.4 |  | . 6 | 0.5 |  | 19.2 | 2.7 | 5.3 | 8.3 | 23.7 | 13.7 |  | 0.5 | 0.1 | 8,724 |
|  | 1995 |  | 4.2 | 17.9 |  | . 5 | 0.3 |  | 24.9 | 4.1 | 4.7 | 6.4 | 12.3 | 18.5 |  | 0.1 | 0.0 | 19,079 |
|  | 1996 |  | 4.0 | 19.0 |  | . 0 | 0.0 |  | 20.0 | 3.0 | 2.0 | 8.0 | 19.0 | 21.0 |  | 0.0 | 0.0 | 39,406 |
|  | 1997 |  | 3.1 | 28.4 |  | . 4 | 0.5 |  | 9.8 | 5.1 | 2.9 | 4.8 | 11.3 | 29.7 |  | 0.0 | 0.1 | 33,556 |
|  | 1998 |  | 2.4 | 31.2 |  | . 0 | 1.3 |  | 6.5 | 6.3 | 5.5 | 3.1 | 8.6 | 29.1 |  | 0.1 | 0.0 | 23,308 |
|  | 1999 |  | 2.7 | 24.0 |  | . 4 | 2.3 |  | 5.4 | 5.2 | 11.8 | 6.4 | 14.5 | 21.9 |  | 0.4 | 0.1 | 19,156 |
|  | 2000 |  | 2.3 | 19.3 |  | . 1 | 0.9 |  | 10.5 | 9.6 | 12.7 | 5.7 | 13.9 | 15.7 |  | 0.1 | 0.1 | 21,793 |
|  | 2001 |  | 3.7 | 28.3 |  | . 4 | 0.2 |  | 15.6 | 6.4 | 6.4 | 5.2 | 9.2 | 17.8 |  | 0.0 | 0.0 | 28,838 |
|  | 2002 |  | 3.8 | 32.9 |  | . 8 | 0.3 |  | 9.9 | 6.7 | 6.6 | 2.5 | 7.2 | 24.4 |  | 0.0 | 0.0 | 38,347 |
|  | 2003 |  | 4.9 | 28.7 |  | . 0 | 1.5 |  | 7.4 | 3.0 | 14.4 | 2.2 | 7.4 | 26.5 |  | 0.0 | 0.0 | 29,382 |
|  | 2004 |  | 4.4 | 31.1 |  | . 1 | 0.5 |  | 6.6 | 1.6 | 12.9 | 2.2 | 11.7 | 26.8 |  | 0.0 | 0.0 | 16,772 |
|  | 2005 |  | 3.7 | 27.5 |  | . 1 | 0.8 |  | 5.4 | 2.4 | 28.1 | 1.7 | 6.4 | 18.8 |  | 0.0 | 0.0 | 15,472 |
|  | 2006 |  | 6.2 | 35.0 |  | . 2 | 0.2 |  | 7.1 | 1.6 | 12.9 | 2.5 | 6.6 | 24.7 |  | 0.0 | 0.0 | 8,636 |
|  | 2007 |  | 5.1 | 28.2 |  | . 6 | 0.3 |  | 6.1 | 1.7 | 17.5 | 1.7 | 4.8 | 32.0 |  | 0.0 | 0.0 | 8,866 |
|  | 2008 |  | 5.1 | 32.7 |  | . 7 | 0.7 |  | 6.4 | 3.2 | 14.6 | 1.0 | 3.1 | 28.6 |  | 0.0 | 0.0 | 7,666 |
|  | 2009 |  | 6.9 | 41.6 |  | . 3 | 0.3 |  | 10.1 | 2.5 | 1.9 | 2.8 | 6.5 | 23.0 |  | 0.0 | 0.0 | 7,146 |
|  | 2010 |  | 6.2 | 31.9 |  | . 7 | 0.0 |  | 12.6 | 1.3 | 1.4 | 3.4 | 9.6 | 30.8 |  | 0.0 | 0.0 | 10,258 |
|  | 2011 |  | 3.6 | 26.5 |  | . 4 | 0.1 |  | 6.7 | 1.3 | 1.4 | 3.1 | 21.9 | 31.9 |  | 0.0 | 0.0 | 9,502 |
|  | 2012 |  | 2.7 | 23.5 |  | . 9 | 0.0 |  | 5.3 | 1.1 | 2.6 | 5.3 | 21.5 | 32.9 |  | 0.0 | 0.0 | 6,378 |
|  | 2013 |  | 4.6 | 26.3 |  | . 3 | 0.2 |  | 8.0 | 2.3 | 2.0 | 4.0 | 15.9 | 30.2 |  | 0.0 | 0.0 | 4,749 |
|  | 2014 |  | 8.7 | 28.0 |  | . 4 | 0.4 |  | 6.4 | 1.2 | 5.2 | 2.5 | 12.3 | 28.7 |  | 0.0 | 0.0 | 5,699 |
|  | 2015 |  | 9.0 | 26.0 |  | . 6 | 0.1 |  | 9.1 | 2.1 | 4.2 | 2.2 | 10.9 | 26.9 |  | 0.0 | 0.0 | 5,890 |
| Average |  |  | 5.0 | 26.1 |  | . 8 | 0.8 |  | 9.0 | 3.9 | 7.5 | 7.4 | 12.4 | 21.6 |  | 0.3 | 0.1 |  |

Table 4.2.4. Faroe Plateau cod (subdivision 5.b.1). Catch in numbers-at-age per fleet in terminal year. Numbers are in thousands and the catch is in tonnes, gutted weight.


Others include gillnetters, industrial bottom trawlers, longlining for halibut, foreign fleets, and scaling to correct catch.
Gutted total catch is calculated as round weight divided by 1.11.

Table 4.2.5. Faroe Plateau cod (subdivision 5.b.1). Number of samples, lengths, otoliths, and individual weights in terminal year.

| Fleet | Size | Samples | Lengths | Otoliths | Weights |
| :--- | :--- | ---: | ---: | ---: | ---: |
| Open boats |  | 4 | 794 | 80 | 794 |
| Longliners | $<100$ GRT | 8 | 1,561 | 160 | 1,561 |
| Longliners | $>100$ GRT | 17 | 3,529 | 358 | 3,529 |
| Jiggers |  | 0 | 0 | 0 | 0 |
| Gillnetters |  | 0 | 0 | 0 | 0 |
| Sing. trawlers | $<400 \mathrm{HP}$ | 0 | 0 | 0 | 0 |
| Sing. trawlers | $400-1000 \mathrm{HP}$ | 11 | 1,981 | 200 | 1,981 |
| Sing. trawlers | $>1000 \mathrm{HP}$ | 0 | 0 | 0 | 0 |
| Pair trawlers | $<1000 \mathrm{HP}$ | 0 | 0 | 0 | 0 |
| Pair trawlers | $>1000 \mathrm{HP}$ | 29 | 5,575 | 439 | 4,412 |
| Total |  | 69 | 13,440 | 1,237 | 12,277 |

Table 4.2.6. Faroe Plateau cod (subdivision 5.b.1). Catch in numbers-at-age used in the XSA model.

| Year \ate | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1959 | 0 | 2002 | 4239 | 858 | 1731 | 200 | 207 | 50 | 10 | 0 |
| 1960 | 0 | 4728 | 4027 | 2574 | 513 | 876 | 171 | 131 | 61 | 0 |
| 1961 | 0 | 3093 | 2686 | 1331 | 1066 | 232 | 372 | 78 | 29 | 0 |
| 1962 | 0 | 4424 | 2500 | 1255 | 855 | 481 | 93 | 94 | 22 | 0 |
| 1963 | 0 | 4110 | 3958 | 1280 | 662 | 284 | 204 | 48 | 30 | 0 |
| 1964 | 0 | 2033 | 3021 | 2300 | 630 | 350 | 158 | 79 | 41 | 0 |
| 1965 | 0 | 852 | 3230 | 2564 | 1416 | 363 | 155 | 48 | 63 | 0 |
| 1966 | 0 | 1337 | 970 | 2080 | 1339 | 606 | 197 | 104 | 33 | 0 |
| 1967 | 0 | 1609 | 2690 | 860 | 1706 | 847 | 309 | 64 | 27 | 0 |
| 1968 | 0 | 1529 | 3322 | 2663 | 945 | 1226 | 452 | 105 | 11 | 0 |
| 1969 | 0 | 878 | 3106 | 3300 | 1538 | 477 | 713 | 203 | 92 | 0 |
| 1970 | 0 | 402 | 1163 | 2172 | 1685 | 752 | 244 | 300 | 44 | 0 |
| 1971 | 0 | 328 | 757 | 821 | 1287 | 1451 | 510 | 114 | 179 | 0 |
| 1972 | 0 |  | 1176 | 810 | 596 | 1021 | 596 | 154 | 25 | 0 |
| 1973 | 0 |  | 3124 | 1590 | 707 | 384 | 312 | 227 | 120 | 97 |
| 1974 | 0 | 2161 | 1266 | 1811 | 934 | 563 | 452 | 149 | 141 | 91 |
| 1975 | 0 | 2584 | 5689 | 2157 | 2211 | 813 | 295 | 190 | 118 | 150 |
| 1976 | 0 | 1497 | 4158 | 3799 | 1380 | 1427 | 617 | 273 | 120 | 186 |
| 1977 | 0 |  | 3282 | 6844 | 3718 | 788 | 1160 | 239 | 134 | 9 |
| 1978 | 0 |  | 1219 | 2643 | 3216 | 1041 | 268 | 201 | 66 | 56 |
| 1979 | 0 | 575 | 1732 | 1673 | 1601 | 1906 | 493 | 134 | 87 | 38 |
| 1980 | 0 | 1129 | 2263 | 1461 | 895 | 807 | 832 | 339 | 42 | 18 |
| 1981 | 0 | 646 | 4137 | 1981 | 947 | 582 | 487 | 527 | 123 | 55 |
| 1982 | 0 | 1139 | 1965 | 3073 | 1286 | 471 | 314 | 169 | 254 | 122 |
| 1983 | 0 | 2149 | 5771 | 2760 | 2746 | 1204 | 510 | 157 | 104 | 102 |
| 1984 | 0 | 4396 | 5234 | 3487 | 1461 | 912 | 314 | 82 | 34 | 66 |
| 1985 | 0 |  | 9484 | 3795 | 1669 | 770 | 872 | 309 | 65 | 80 |
| 1986 | 0 | 210 | 3586 | 8462 | 2373 | 907 | 236 | 147 | 47 | 38 |
| 1987 | 0 | 257 | 1362 | 2611 | 3083 | 812 | 224 | 68 | 69 | 26 |
| 1988 | 0 | 509 | 2122 | 1945 | 1484 | 2178 | 492 | 168 | 33 | 25 |
| 1989 | 0 | 2237 | 2151 | 2187 | 1121 | 1026 | 997 | 220 | 61 | 9 |
| 1990 | 0 | 247 | 2892 | 1504 | 865 | 410 | 298 | 295 | 51 | 26 |
| 1991 | 0 | 192 | 451 | 2152 | 622 | 303 | 142 | 93 | 53 | 24 |
| 1992 | 0 | 205 | 455 | 466 | 911 | 293 | 132 | 53 | 30 | 34 |
| 1993 | 0 |  | 802 | 603 | 222 | 329 | 96 | 33 | 22 | 25 |
| 1994 | 0 | 573 | 788 | 1062 | 532 | 125 | 176 | 39 | 23 | 16 |
| 1995 | 0 | 2615 | 2716 | 2008 | 1012 | 465 | 118 | 175 | 44 | 49 |
| 1996 | 0 | 351 | 5164 | 4608 | 1542 | 1526 | 596 | 147 | 347 | 47 |
| 1997 | 0 | 200 | 1278 | 6710 | 3731 | 657 | 639 | 170 | 51 | 120 |
| 1998 | 0 |  | 745 | 1558 | 5140 | 1529 | 159 | 118 | 28 | 25 |
| 1999 | 0 | 1246 | 1044 | 840 | 1164 | 2339 | 461 | 62 | 18 | 8 |


| Year\AGE | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2000 | 0 | 2170 | 2737 | 811 | 443 | 700 | 840 | 108 | 8 | 1 |
| 2001 | 0 | 3967 | 3812 | 2130 | 373 | 372 | 728 | 443 | 36 | 6 |
| 2002 | 0 | 2099 | 7354 | 3405 | 1688 | 474 | 538 | 417 | 293 | 7 |
| 2003 | 0 | 697 | 2186 | 4696 | 1979 | 657 | 182 | 94 | 118 | 21 |
| 2004 | 0 | 98 | 673 | 1230 | 2051 | 717 | 234 | 63 | 41 | 36 |
| 2005 | 0 | 504 | 604 | 896 | 1146 | 841 | 208 | 41 | 19 | 31 |
| 2006 | 0 | 1110 | 1097 | 469 | 663 | 801 | 333 | 76 | 10 | 3 |
| 2007 | 0 | 506 | 1226 | 723 | 315 | 289 | 255 | 85 | 20 | 3 |
| 2008 | 0 | 287 | 761 | 783 | 430 | 187 | 157 | 156 | 57 | 19 |
| 2009 | 0 | 873 | 2262 | 861 | 618 | 296 | 85 | 55 | 43 | 17 |
| 2010 | 0 | 2113 | 2034 | 861 | 468 | 481 | 178 | 58 | 33 | 38 |
| 2011 | 0 | 328 | 2343 | 1234 | 365 | 188 | 126 | 50 | 19 | 2 |
| 2012 | 0 | 49 | 517 | 1346 | 555 | 200 | 99 | 69 | 25 | 22 |
| 2013 | 0 | 55 | 173 | 333 | 587 | 175 | 39 | 25 | 15 | 5 |
| 2014 | 0 | 387 | 518 | 286 | 499 | 350 | 86 | 14 | 9 | 1 |
| 2015 | 0 | 156 | 1035 | 522 | 210 | 282 | 221 | 47 | 23 | 7 |

Table 4.2.7. Faroe Plateau cod (subdivision 5.b.1). Mean weight at age (kg) in the catches.

| Year \age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1959 | 0.850 | 1.730 | 3.230 | 4.400 | 5.800 | 6.370 | 7.340 | 7.880 | 10.270 |
| 1960 | 1.000 | 2.030 | 3.370 | 4.420 | 6.020 | 6.650 | 8.120 | 11.000 | 10.270 |
| 1961 | 1.080 | 2.220 | 3.450 | 4.690 | 5.520 | 7.090 | 9.910 | 8.030 | 10.270 |
| 1962 | 1.000 | 2.270 | 3.350 | 4.580 | 4.930 | 9.080 | 6.590 | 6.660 | 10.270 |
| 1963 | 1.040 | 1.940 | 3.510 | 4.600 | 5.500 | 6.780 | 8.710 | 11.720 | 10.820 |
| 1964 | 0.970 | 1.830 | 3.150 | 4.330 | 6.080 | 7.000 | 6.250 | 6.190 | 14.390 |
| 1965 | 0.920 | 1.450 | 2.570 | 3.780 | 5.690 | 7.310 | 7.930 | 8.090 | 11.110 |
| 1966 | 0.980 | 1.770 | 2.750 | 3.510 | 4.800 | 6.320 | 7.510 | 10.340 | 11.650 |
| 1967 | 0.960 | 1.930 | 3.130 | 4.040 | 4.780 | 6.250 | 7.000 | 11.010 | 10.690 |
| 1968 | 0.880 | 1.720 | 3.070 | 4.120 | 4.650 | 5.500 | 7.670 | 10.950 | 9.280 |
| 1969 | 1.090 | 1.800 | 2.850 | 3.670 | 4.890 | 5.050 | 7.410 | 8.660 | 14.390 |
| 1970 | 0.960 | 2.230 | 2.690 | 3.940 | 5.140 | 6.460 | 10.310 | 7.390 | 9.340 |
| 1971 | 0.810 | 1.800 | 2.980 | 3.580 | 3.940 | 4.870 | 6.480 | 6.370 | 10.220 |
| 1972 | 0.660 | 1.610 | 2.580 | 3.260 | 4.290 | 4.950 | 6.480 | 6.900 | 11.550 |
| 1973 | 1.110 | 2.000 | 3.410 | 3.890 | 5.100 | 5.100 | 6.120 | 8.660 | 7.570 |
| 1974 | 1.080 | 2.220 | 3.440 | 4.800 | 5.180 | 5.880 | 6.140 | 8.630 | 7.620 |
| 1975 | 0.790 | 1.790 | 2.980 | 4.260 | 5.460 | 6.250 | 7.510 | 7.390 | 8.170 |
| 1976 | 0.940 | 1.720 | 2.840 | 3.700 | 5.260 | 6.430 | 6.390 | 8.550 | 13.620 |
| 1977 | 0.870 | 1.790 | 2.530 | 3.680 | 4.650 | 5.340 | 6.230 | 8.380 | 10.720 |
| 1978 | 1.112 | 1.385 | 2.140 | 3.125 | 4.363 | 5.927 | 6.348 | 8.715 | 12.229 |
| 1979 | 0.897 | 1.682 | 2.211 | 3.052 | 3.642 | 4.719 | 7.272 | 8.368 | 13.042 |
| 1980 | 0.927 | 1.432 | 2.220 | 3.105 | 3.539 | 4.392 | 6.100 | 7.603 | 9.668 |
| 1981 | 1.080 | 1.470 | 2.180 | 3.210 | 3.700 | 4.240 | 4.430 | 6.690 | 10.000 |


| Year \age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 1.230 | 1.413 | 2.138 | 3.107 | 4.012 | 5.442 | 5.563 | 5.216 | 6.707 |
| 1983 | 1.338 | 1.950 | 2.403 | 3.107 | 4.110 | 5.020 | 5.601 | 8.013 | 8.031 |
| 1984 | 1.195 | 1.888 | 2.980 | 3.679 | 4.470 | 5.488 | 6.466 | 6.628 | 10.981 |
| 1985 | 0.905 | 1.658 | 2.626 | 3.400 | 3.752 | 4.220 | 4.739 | 6.511 | 10.981 |
| 1986 | 1.099 | 1.459 | 2.046 | 2.936 | 3.786 | 4.699 | 5.893 | 9.700 | 8.815 |
| 1987 | 1.093 | 1.517 | 2.160 | 2.766 | 3.908 | 5.461 | 6.341 | 8.509 | 9.811 |
| 1988 | 1.061 | 1.749 | 2.300 | 2.914 | 3.109 | 3.976 | 4.896 | 7.087 | 8.287 |
| 1989 | 1.010 | 1.597 | 2.200 | 2.934 | 3.468 | 3.750 | 4.682 | 6.140 | 9.156 |
| 1990 | 0.945 | 1.300 | 1.959 | 2.531 | 3.273 | 4.652 | 4.758 | 6.704 | 8.689 |
| 1991 | 0.779 | 1.271 | 1.570 | 2.524 | 3.185 | 4.086 | 5.656 | 5.973 | 8.147 |
| 1992 | 0.989 | 1.364 | 1.779 | 2.312 | 3.477 | 4.545 | 6.275 | 7.619 | 9.725 |
| 1993 | 1.155 | 1.704 | 2.421 | 3.132 | 3.723 | 4.971 | 6.159 | 7.614 | 9.587 |
| 1994 | 1.194 | 1.843 | 2.613 | 3.654 | 4.584 | 4.976 | 7.146 | 8.564 | 8.796 |
| 1995 | 1.218 | 1.986 | 2.622 | 3.925 | 5.180 | 6.079 | 6.241 | 7.782 | 8.627 |
| 1996 | 1.016 | 1.737 | 2.745 | 3.800 | 4.455 | 4.978 | 5.270 | 5.593 | 7.482 |
| 1997 | 0.901 | 1.341 | 1.958 | 3.012 | 4.158 | 4.491 | 5.312 | 6.172 | 7.056 |
| 1998 | 1.004 | 1.417 | 1.802 | 2.280 | 3.478 | 5.433 | 5.851 | 7.970 | 8.802 |
| 1999 | 1.050 | 1.586 | 2.350 | 2.774 | 3.214 | 5.496 | 8.276 | 9.129 | 10.652 |
| 2000 | 1.416 | 2.170 | 3.187 | 3.795 | 4.048 | 4.577 | 8.182 | 11.895 | 13.009 |
| 2001 | 1.164 | 2.076 | 3.053 | 3.976 | 4.394 | 4.871 | 5.563 | 7.277 | 12.394 |
| 2002 | 1.017 | 1.768 | 2.805 | 3.529 | 4.095 | 4.475 | 4.650 | 6.244 | 7.457 |
| 2003 | 0.820 | 1.362 | 2.127 | 3.329 | 4.092 | 4.670 | 6.000 | 6.727 | 6.810 |
| 2004 | 1.037 | 1.154 | 1.693 | 2.363 | 3.830 | 5.191 | 6.326 | 7.656 | 9.573 |
| 2005 | 0.986 | 1.373 | 1.760 | 2.293 | 3.138 | 5.287 | 8.285 | 8.703 | 9.517 |
| 2006 | 0.839 | 1.304 | 1.988 | 2.386 | 3.330 | 4.691 | 7.635 | 9.524 | 11.990 |
| 2007 | 0.937 | 1.324 | 1.970 | 3.076 | 3.529 | 4.710 | 6.464 | 9.461 | 9.509 |
| 2008 | 1.209 | 1.478 | 2.104 | 2.714 | 3.804 | 4.669 | 5.915 | 7.233 | 9.559 |
| 2009 | 0.805 | 1.431 | 2.287 | 2.723 | 3.435 | 5.081 | 6.281 | 8.312 | 9.959 |
| 2010 | 1.049 | 1.642 | 2.400 | 3.212 | 3.678 | 4.774 | 5.973 | 7.094 | 9.800 |
| 2011 | 0.815 | 1.367 | 2.413 | 3.493 | 4.525 | 5.076 | 6.631 | 6.863 | 10.089 |
| 2012 | 1.007 | 1.315 | 1.893 | 3.102 | 4.279 | 5.573 | 5.871 | 7.482 | 9.206 |
| 2013 | 1.011 | 1.527 | 2.528 | 3.180 | 4.672 | 6.776 | 6.966 | 9.028 | 10.324 |
| 2014 | 1.099 | 1.653 | 2.466 | 3.000 | 4.148 | 6.489 | 9.394 | 9.236 | 12.120 |
| 2015 | 1.198 | 1.733 | 2.769 | 3.650 | 4.403 | 5.768 | 8.035 | 10.334 | 11.127 |
| 2016 | 1.057 | 1.857 | 2.706 | 3.686 | 4.237 | 5.057 | 6.472 | 9.644 | 9.644 |

Table 4.2.8. Faroe Plateau cod (subdivision 5.b.1). Proportion mature at age. From 1961-1982 the average from 1983-1996 is used (as it was used in the 1990s). In 2002, the high maturities for age 2 in 1983 (0.63), 1984 (0.4) and in 1993 (0.25) were revised, but not the maturities back in time.

| Year/ | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1959 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1960 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1961 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1962 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1963 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1964 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1965 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1966 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1967 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1968 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1969 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1970 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1971 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1972 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1973 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1974 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1975 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1976 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1977 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1978 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1979 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1980 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1981 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1982 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1983 | 0.00 | 0.03 | 0.71 | 0.93 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1984 | 0.00 | 0.07 | 0.96 | 0.98 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1985 | 0.00 | 0.00 | 0.50 | 0.96 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1986 | 0.00 | 0.00 | 0.38 | 0.93 | 1.00 | 1.00 | 0.96 | 0.94 | 1.00 | 1.00 |
| 1987 | 0.00 | 0.00 | 0.67 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1988 | 0.00 | 0.06 | 0.72 | 0.90 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1989 | 0.00 | 0.05 | 0.54 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1990 | 0.00 | 0.00 | 0.68 | 0.90 | 0.99 | 0.96 | 0.98 | 1.00 | 1.00 | 1.00 |
| 1991 | 0.00 | 0.00 | 0.72 | 0.86 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1992 | 0.00 | 0.06 | 0.50 | 0.82 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1993 | 0.00 | 0.03 | 0.73 | 0.78 | 0.91 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1994 | 0.00 | 0.05 | 0.33 | 0.88 | 0.96 | 1.00 | 0.96 | 1.00 | 1.00 | 1.00 |
| 1995 | 0.00 | 0.09 | 0.35 | 0.33 | 0.66 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1996 | 0.00 | 0.04 | 0.43 | 0.74 | 0.85 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1997 | 0.00 | 0.00 | 0.64 | 0.91 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1998 | 0.00 | 0.00 | 0.62 | 0.90 | 0.99 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1999 | 0.00 | 0.02 | 0.43 | 0.88 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2000 | 0.00 | 0.02 | 0.39 | 0.69 | 0.92 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |


| YEAR/AGE | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2001 | 0.00 | 0.07 | 0.47 | 0.86 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2002 | 0.00 | 0.04 | 0.37 | 0.76 | 0.97 | 0.93 | 0.97 | 1.00 | 1.00 | 1.00 |
| 2003 | 0.00 | 0.00 | 0.29 | 0.79 | 0.88 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2004 | 0.00 | 0.00 | 0.51 | 0.78 | 0.92 | 0.89 | 0.87 | 1.00 | 1.00 | 1.00 |
| 2005 | 0.00 | 0.05 | 0.66 | 0.90 | 0.93 | 0.98 | 0.92 | 1.00 | 1.00 | 1.00 |
| 2006 | 0.00 | 0.04 | 0.59 | 0.80 | 0.99 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2007 | 0.00 | 0.00 | 0.47 | 0.78 | 0.91 | 0.99 | 0.97 | 1.00 | 1.00 | 1.00 |
| 2008 | 0.00 | 0.10 | 0.78 | 0.91 | 0.90 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2009 | 0.00 | 0.09 | 0.61 | 0.81 | 0.96 | 0.94 | 0.96 | 1.00 | 1.00 | 1.00 |
| 2010 | 0.00 | 0.08 | 0.61 | 0.77 | 0.94 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2011 | 0.00 | 0.06 | 0.51 | 0.69 | 0.84 | 0.93 | 0.98 | 1.00 | 1.00 | 1.00 |
| 2012 | 0.00 | 0.00 | 0.63 | 0.85 | 0.94 | 0.97 | 1.00 | 1.00 | 1.00 | 0.83 |
| 2013 | 0.00 | 0.24 | 0.82 | 0.95 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2014 | 0.00 | 0.24 | 0.73 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2015 | 0.00 | 0.28 | 0.48 | 0.70 | 0.95 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |

Table 4.2.9. Faroe Plateau cod (subdivision 5.b.1). Summer survey tuning series (number of individuals per 200 stations) and spring survey tuning series (number of individuals per $\mathbf{1 0 0}$ stations) used as tuning series in the XSA model.

| FAROE PLATEAU <br> vised.TXT | COD ( | SUBDIVISION 5.B.1) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 102 |  |  |  |  |  |  |
| SUMMER SURVEY |  |  |  |  |  |  |
| 19962015 |  |  |  |  |  |  |
| 110.60 .7 |  |  |  |  |  |  |
| 28 |  |  |  |  |  |  |
| 200707 | 6576.5 | 3705.1 | 1298.1 | 701.5 | 233.1 | 48.5 |
| 200512.7 | 1500.7 | 6754.6 | 1466.6 | 178.4 | 137.8 | 30.1 |
| 200524.9 | 505.1 | 979.4 | 3675.2 | 902.6 | 50 | 37 |
| 200373.3 | 1256.8 | 753.1 | 675.3 | 1422.5 | 238 | 40.4 |
| 2001364.1 | 1153.3 | 673.8 | 309.6 | 436.9 | 600.8 | 35.4 |
| 2003422.1 | 2458.7 | 1537.8 | 415.9 | 234.8 | 283 | 242 |
| 2002326 | 5562.9 | 1816.5 | 810.8 | 147.7 | 83.3 | 69.5 |
| 200354 | 1038.8 | 2209.2 | 565.9 | 123.4 | 17.6 | 11.9 |
| 200437 | 839.9 | 1080.2 | 1550.2 | 344.2 | 80.2 | 25.7 |
| 200616.5 | 735.1 | 872.1 | 1166.3 | 756 | 142.5 | 44.8 |
| 200978.4 | 684.2 | 349.3 | 312 | 256.6 | 123 | 28.2 |
| 200234.1 | 448.7 | 314.2 | 179.7 | 134.5 | 75.9 | 30.9 |
| 20068.8 | 370.1 | 328 | 401.2 | 160.1 | 52.4 | 27.5 |
| 200428.2 | 1980.6 | 817.7 | 551.4 | 393.1 | 132.1 | 47.8 |
| 2001239.3 | 1543.9 | 1012 | 363.4 | 243.6 | 148.9 | 41.5 |
| 200301.7 | 1373.6 | 1084.2 | 380.1 | 160.6 | 104.6 | 37.4 |
| 20022.1 | 230.8 | 1081.8 | 511.7 | 88.4 | 35.8 | 19.5 |
| 200101.7 | 205.9 | 209.3 | 888.4 | 542.5 | 104.2 | 43.9 |
| 200642.3 | 861.2 | 357.6 | 358.2 | 401.5 | 124.3 | 36.6 |
| 200235.3 | 2230.4 | 1696.1 | 414.7 | 363.4 | 242.3 | 67.2 |


| SPRING SURVEY (shifted back to december) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{array}{lllll}1 & 1 & 0.9 & 1.0\end{array}$ |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |
| 100 | 612.5 | 336.9 | 912.8 | 508.5 | 129.7 | 187.2 |
| 28.6 |  |  |  |  |  |  |
| 100 | 623.2 | 845.7 | 1528.4 | 1525.2 | 1191.4 | 285.6 |
| 350.8 48.9 |  |  |  |  |  |  |
| 100 | 215.5 | 4043.9 | 3984.4 | 1892.1 | 1372 | 420.8 |
| 82.8169 .7 |  |  |  |  |  |  |
| 100 | 72.5 | 834.4 | 5398.3 | 2359.5 | 333.9 | 227 |
| 58.8 5.3 |  |  |  |  |  |  |
| 100 | 69.7 | 425.2 | 1572.1 | 4919.3 | 1136 | 82.3 |
| $40.7 \quad 35.2$ |  |  |  |  |  |  |
| 100 | 704.7 | 674.9 | 991.3 | 1225.2 | 2079.2 | 252.1 |
| 25.213 .4 |  |  |  |  |  |  |
| 100 | 316 | 1432.4 | 746.1 | 441 | 506.7 | 836.7 |
| 63.8 3.1 |  |  |  |  |  |  |
| 100 | 938.4 | 2387.8 | 1993.8 | 456.2 | 324.4 | 578.6 |
| 128.6 3.9 |  |  |  |  |  |  |
| 100 | 383 | 4564.1 | 2892.1 | 1579.7 | 331.9 | 231.8 |
| $178.9 \quad 131.9$ |  |  |  |  |  |  |
| 100 | 90.2 | 719 | 3915 | 1260.4 | 528.7 | 67.4 |
| $51.7 \quad 39.7$ |  |  |  |  |  |  |
| 100 | 609.5 | 575.8 | 844.6 | 1175.1 | 292.9 | 66 |
| 22.2 |  |  |  |  |  |  |
| 100 | 383.1 | 438.2 | 1151.7 | 1440.2 | 844.5 | 140.6 |
| 14 3.8 |  |  |  |  |  |  |
| 100 | 167.5 | 156.7 | 177.3 | 360.1 | 292 | 95 |
| 15.5 4 |  |  |  |  |  |  |
| 100 | 41.1 | 270.9 | 286.6 | 155.2 | 170.4 | 105.1 |
| 37.814 .4 |  |  |  |  |  |  |
| 100 | 176.6 | 474.5 | 851.9 | 479.2 | 151.5 | 83.9 |
| 39.413 .3 |  |  |  |  |  |  |
| 100 | 307.8 | 475.5 | 977.7 | 1159.1 | 427.3 | 73.7 |
| $31.6 \quad 24.9$ |  |  |  |  |  |  |
| 100 | 697.6 | 1318.8 | 745.6 | 538.1 | 381 | 98.9 |
| $41 \quad 17.2$ |  |  |  |  |  |  |
| 100 | 148.4 | 1319 | 1240.3 | 562.4 | 300.2 | 237.8 |
| 85.221 .9 |  |  |  |  |  |  |
| 100 | 41.1 | 273.8 | 1303.8 | 326.7 | 73.6 | 27 |
| 23.7 |  |  |  |  |  |  |
| 100 | 68 | 377.6 | 1699.8 | 2053.2 | 295.6 | 32.6 |
| 22.417 .7 |  |  |  |  |  |  |
| 100 | 130.9 | 113.4 | 159.6 | 419.7 | 333 | 74.8 |
| $22 \quad 13.6$ |  |  |  |  |  |  |
| 100 | 22.4 | 533.3 | 225.6 | 193.9 | 305.2 | 138.9 |
| 32.6 | 8 |  |  |  |  |  |
| 100 | 81.7 | 280.1 | 697.3 | 151.8 | 73.4 | 77.3 |
| 27.2 7.7 |  |  |  |  |  |  |

Table 4.2.10. Faroe Plateau cod (subdivision 5.b.1). Pairtrawler abundance index (number of individuals per 1000 fishing hours). This series was not used in the tuning of the XSA. The season is June-December. The otoliths are selected from deep (> $\mathbf{1 5 0} \mathbf{~ m}$ ) locations.

| Year \age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 1200 | 1638 | 1783 | 1381 | 928 | 719 | 297 | 194 |
| 1990 | 116 | 2856 | 2057 | 834 | 465 | 419 | 200 | 0 |
| 1991 | 8 | 148 | 1401 | 869 | 329 | 225 | 65 | 93 |
| 1992 | 84 | 487 | 696 | 1234 | 760 | 353 | 129 | 62 |
| 1993 | 51 | 1081 | 2192 | 746 | 1062 | 398 | 67 | 107 |
| 1994 | 1314 | 2129 | 1457 | 2208 | 697 | 1241 | 461 | 53 |
| 1995 | 577 | 3645 | 5178 | 4199 | 2769 | 543 | 539 | 106 |
| 1996 | 242 | 10608 | 16683 | 7985 | 4410 | 194 | 0 | 723 |
| 1997 | 28 | 674 | 6038 | 9375 | 2413 | 944 | 113 | 0 |
| 1998 | 80 | 731 | 1805 | 5941 | 4904 | 801 | 286 | 0 |
| 1999 | 444 | 2082 | 1933 | 3008 | 5136 | 2220 | 218 | 4 |
| 2000 | 3478 | 3956 | 1737 | 956 | 1003 | 1694 | 382 | 0 |
| 2001 | 3385 | 6700 | 3009 | 555 | 415 | 797 | 862 | 25 |
| 2002 | 571 | 6409 | 5019 | 1235 | 432 | 400 | 41 | 228 |
| 2003 | 63 | 1341 | 4450 | 3630 | 870 | 270 | 152 | 145 |
| 2004 | 23 | 0 | 278 | 2534 | 2831 | 1733 | 274 | 184 |
| 2005 | 42 | 399 | 655 | 1766 | 2171 | 860 | 148 | 70 |
| 2006 | 93 | 135 | 699 | 755 | 1580 | 612 | 787 | 71 |
| 2007 | 64 | 916 | 1767 | 1392 | 802 | 656 | 206 | 46 |
| 2008 | 54 | 295 | 418 | 573 | 387 | 456 | 487 | 182 |
| 2009 | 11 | 734 | 801 | 756 | 448 | 247 | 147 | 105 |
| 2010 | 1578 | 2917 | 1787 | 543 | 603 | 190 | 0 | 81 |
| 2011 | 22 | 1487 | 4078 | 1967 | 622 | 441 | 95 | 25 |
| 2012 | 0 | 95 | 1531 | 1789 | 950 | 223 | 40 | 107 |
| 2013 | 35 | 102 | 761 | 1583 | 670 | 103 | 57 | 36 |
| 2014 | 292 | 1631 | 1006 | 1690 | 1812 | 477 | 94 | 101 |
| 2015 | 43 | 967 | 1943 | 1019 | 1190 | 1086 | 320 | 96 |

Table 4.2.11. Faroe Plateau cod (subdivision 5.b.1). Longliner abundance index (number of individuals per 100000 hooks). This series was not used in the tuning of the XSA. The age composition was obtained from all longliners > 100 GRT. The area was restricted to the area west of Faroe Islands at depths between 100 and 200 m .

| Year \aGe | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 405 | 2610 | 9306 | 3330 | 806 | 2754 | 847 | 258 |
| 1994 | 101 | 8105 | 14105 | 7863 | 4659 | 962 | 1187 | 71 |
| 1995 | 0 | 15249 | 23062 | 2895 | 2505 | 1568 | 708 | 1073 |
| 1996 | 0 | 2269 | 18658 | 13265 | 4153 | 8435 | 4513 | 1147 |
| 1997 | 0 | 1738 | 5837 | 26368 | 18089 | 2805 | 2807 | 402 |
| 1998 | 1892 | 4490 | 2025 | 2565 | 11738 | 2732 | 131 | 19 |
| 1999 | 849 | 10968 | 3811 | 985 | 1891 | 3759 | 548 | 109 |
| 2000 | 2695 | 10983 | 6710 | 998 | 780 | 1473 | 2136 | 109 |
| 2001 | 287 | 12999 | 7409 | 2660 | 515 | 1135 | 1808 | 2545 |
| 2002 | 105 | 6862 | 20902 | 10819 | 7759 | 1561 | 1945 | 1265 |
| 2003 | 16 | 2099 | 6057 | 15910 | 7778 | 1830 | 708 | 650 |
| 2004 | 59 | 510 | 1773 | 2438 | 3214 | 1059 | 293 | 71 |
| 2005 | 297 | 2169 | 1543 | 2313 | 2327 | 1360 | 170 | 13 |
| 2006 | 151 | 5813 | 5319 | 674 | 2205 | 2352 | 1148 | 56 |
| 2007 | 274 | 3578 | 6383 | 2778 | 1927 | 1159 | 1118 | 134 |
| 2008 | 1270 | 2243 | 4449 | 4773 | 2564 | 1133 | 816 | 716 |
| 2009 | 294 | 2670 | 15107 | 6308 | 3028 | 2491 | 683 | 132 |
| 2010 | 23 | 20287 | 16914 | 8733 | 2595 | 4780 | 1878 | 864 |
| 2011 | 160 | 2817 | 28218 | 14391 | 4295 | 2207 | 1252 | 195 |
| 2012 | 0 | 1833 | 9562 | 8309 | 2364 | 1296 | 403 | 197 |
| 2013 | 0 | 52 | 209 | 2887 | 5132 | 2654 | 1222 | 359 |
| 2014 | 93 | 5898 | 9602 | 4695 | 4398 | 3475 | 1289 | 116 |
| 2015 | 0 | 1260 | 10417 | 8202 | 3167 | 3342 | 2428 | 414 |

Table 4.2.12. Longliner abundance index (number of individuals per day) for longliners < 25 GRT operating mainly near shore. This series was not used in the tuning of the XSA. The age composition was obtained from all longliners.

| YeAR\AG |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | 0.9 |  | 7.5 | 4.7 | 3.8 | 1.6 | 0.9 | 0.5 | 0.2 |
| 1984 | 0.0 |  | 33.3 | 32.1 | 13.2 | 5.8 | 6.3 | 1.0 | 0.7 |
| 1985 | 0.0 |  | 3.7 | 50.1 | 35.0 | 25.3 | 14.1 | 19.6 | 5.8 |
| 1986 | 0.0 |  | 5.6 | 41.6 | 24.0 | 15.3 | 6.8 | 6.2 | 2.2 |
| 1987 | 0.0 |  | 6.8 | 11.3 | 16.6 | 27.5 | 12.4 | 5.3 | 0.9 |
| 1988 | 0.0 |  | 3.1 | 6.4 | 13.0 | 8.5 | 19.1 | 6.5 | 2.6 |
| 1989 | 0.1 |  | 43.7 | 21.3 | 20.5 | 13.9 | 7.5 | 16.1 | 2.2 |
| 1990 | 0.0 |  | 7.9 | 40.3 | 8.6 | 12.2 | 6.5 | 7.7 | 4.2 |
| 1991 | 0.0 |  | 0.0 | 5.2 | 27.0 | 8.7 | 3.9 | 2.4 | 0.7 |
| 1992 | 0.0 |  | 6.2 | 17.1 | 6.9 | 3.9 | 3.6 | 1.8 | 1.4 |
| 1993 | 0.4 |  | 4.6 | 19.2 | 7.3 | 1.4 | 1.3 | 0.3 | 1.3 |
| 1994 | 0.1 |  | 14.9 | 18.4 | 15.4 | 6.6 | 2.1 | 2.6 | 0.5 |
| 1995 | 0.0 |  | 53.6 | 47.8 | 12.2 | 8.4 | 5.1 | 2.0 | 3.1 |
| 1996 | 0.0 |  | 5.9 | 76.2 | 52.1 | 13.1 | 28.8 | 14.3 | 4.2 |
| 1997 | 0.0 |  | 4.6 | 16.6 | 71.8 | 54.5 | 7.9 | 7.6 | 0.9 |
| 1998 | 5.8 |  | 12.1 | 5.6 | 8.2 | 33.1 | 9.9 | 0.4 | 0.4 |
| 1999 | 0.3 |  | 29.2 | 10.0 | 4.7 | 7.0 | 15.9 | 2.5 | 0.1 |
| 2000 | 9.6 |  | 40.4 | 23.5 | 1.3 | 1.3 | 2.4 | 4.2 | 0.5 |
| 2001 | 0.6 |  | 96.6 | 48.7 | 17.1 | 3.0 | 5.7 | 12.6 | 12.9 |
| 2002 | 0.1 |  | 47.6 | 97.2 | 43.4 | 30.0 | 7.3 | 11.5 | 6.8 |
| 2003 | 0.0 |  | 17.5 | 37.4 | 106.4 | 59.1 | 12.9 | 4.1 | 1.5 |
| 2004 | 0.0 |  | 7.0 | 21.5 | 21.0 | 31.1 | 8.2 | 0.3 | 0.0 |
| 2005 | 0.6 |  | 14.7 | 20.5 | 18.5 | 32.9 | 15.6 | 1.5 | 0.0 |
| 2006 | 2.0 |  | 58.7 | 47.0 | 9.1 | 10.6 | 13.6 | 4.1 | 0.4 |
| 2007 | 0.2 |  | 11.2 | 23.2 | 8.9 | 4.2 | 4.9 | 3.5 | 0.6 |
| 2008 | 0.3 |  | 3.4 | 16.2 | 21.1 | 14.4 | 3.3 | 1.5 | 2.1 |
| 2009 | 3.1 |  | 33.3 | 154.6 | 57.5 | 33.9 | 23.5 | 9.6 | 5.9 |
| 2010 | 2.6 |  | 135.7 | 147.1 | 62.4 | 27.3 | 28.5 | 8.5 | 1.8 |
| 2011 | 0.0 |  | 19.7 | 156.5 | 65.0 | 25.2 | 15.6 | 8.5 | 1.9 |
| 2012 | 0.3 |  | 4.6 | 39.3 | 59.0 | 15.1 | 5.2 | 2.6 | 1.3 |
| 2013 | 1.2 |  | 16.6 | 23.8 | 63.6 | 58.0 | 7.8 | 2.9 | 0.0 |
| 2014 | 2.1 |  | 103.4 | 102.0 | 46.9 | 27.3 | 17.1 | 1.4 | 0.0 |
| 2015 | 0.9 |  | 25.4 | 148.6 | 65.3 | 23.0 | 17.9 | 10.7 | 0.7 |

Table 4.6.1. Faroe Plateau cod (subdivision 5.b.1). The XSA-run.

```
Lowestoft VPA Version 3.1
    12/04/2016 8:57
Extended Survivors Analysis
COD FAROE PLATEAU (ICES SUBDIVISION 5.b.1) COD_ind_Surveys_re-
vised
cpue data from file Surveys_revised_1replacedvalue.TXT
Catch data for 57 years. 1959 to 2015. Ages 1 to 10.
    Fleet, First, Last, First, Last, Alpha, Beta
SUMMER SURVEY ', year, year, age , age . 1996, 2015, 2, 8, .600, . 700
SPRING SURVEY (shift, 1993, 2015, 1, 8, .900, 1.000
Time-series weights :
    Tapered time weighting not applied
Catchability analysis :
    Catchability independent of stock size for all ages
    Catchability independent of age for ages >= 6
Terminal population estimation :
```

    Survivor estimates shrunk towards the mean F
    of the final 5 years or the 5 oldest ages.
    S.E. of the mean to which the estimates are shrunk \(=2.000\)
    Minimum standard error for population
    estimates derived from each fleet \(=.300\)
    Prior weighting not applied
    Tuning converged after 29 iterations
Regression weights
, $1.000,1.000,1.000,1.000,1.000,1.000,1.000,1.000,1.000,1.000$
Fishing mortalities
Age, 2006, 2007, 2008, 2009, 2010, 2011, 2012, 2013, 2014, 2015

| 1, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2, | .181, | .118, | .050, | .125, | .231, | .088, | .029, | .021, | .057, | .075 |
| 3, | .326, | .313, | .261, | .687, | .474, | .435, | .195, | .137, | .284, | .215 |
| 4, | .352, | .371, | .337, | .531, | .614, | .597, | .482, | .186, | .350, | .518 |
| 5, | .601, | .426, | .396, | .489, | .627, | .578, | .595, | .400, | .469, | .472 |
| 6, | .808, | .577, | .486, | .524, | .914, | .558, | .741, | .376, | .443, | .533 |
| 7, | .946, | .661, | .730, | .426, | .706, | .650, | .655, | .303, | .320, | .562 |
| 8, | 1.004, | .675, | 1.204, | .616, | .585, | .434, | .946, | .336, | .169, | .290 |
| 9, | .292, | .811, | 1.562, | 1.542, | .978, | .383, | .403, | .542, | .193, | .461 |

XSA population numbers (Thousands)

| AGE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 8, | 9, 1' | 2, | 3 , | , | 5, | 6, | , |
| 2006 | 6.14 E | 39 E | 36 E | 75E | 62E | 60E |  |
| $6.02 \mathrm{E}+02,1.33 \mathrm{E}+02,4.36 \mathrm{E}+01$, |  |  |  |  |  |  |  |
| 2007 | 7.90E+03, 5.03E+03, |  |  |  |  |  |  |
| 5.83E+02, | $1.91 \mathrm{E}+$ | E+ |  |  |  |  |  |

```
2008, 1.01E+04, 6.46E+03, 3.66E+03, 3.02E+03, 1.45E+03, 5.37E+02,
3.35E+02, 2.46E+02, 7.97E+01,
2009, 1.38E+04, 8.24E+03, 5.03E+03, 2.31E+03, 1.77E+03, 8.02E+02,
2.71E+02, 1.32E+02, 6.05E+01,
    2010, 5.24E+03, 1.13E+04, 5.95E+03, 2.07E+03, 1.11E+03, 8.88E+02,
3.89E+02, 1.45E+02, 5.84E+01,
2011, 2.30E+03, 4.29E+03, 7.34E+03, 3.03E+03, 9.19E+02, 4.86E+02,
2.91E+02, 1.57E+02, 6.60E+01,
    2012, 3.53E+03, 1.88E+03, 3.22E+03, 3.89E+03, 1.37E+03, 4.22E+02,
2.28E+02, 1.25E+02, 8.34E+01,
2013, 9.35E+03, 2.89E+03, 1.50E+03, 2.17E+03, 1.97E+03, 6.17E+02,
1.65E+02, 9.68E+01, 3.96E+01,
    2014, 2.92E+03, 7.65E+03, 2.32E+03, 1.07E+03, 1.47E+03, 1.08E+03,
3.47E+02, 9.95E+01, 5.66E+01,
    2015, 4.12E+03, 2.39E+03, 5.92E+03, 1.43E+03, 6.17E+02, 7.55E+02,
    5.68E+02, 2.06E+02, 6.88E+01,
```

    Estimated population abundance at 1st Jan 2016
    \(0.00 \mathrm{E}+00,3.37 \mathrm{E}+03,1.81 \mathrm{E}+03,3.91 \mathrm{E}+03,6.97 \mathrm{E}+02,3.15 \mathrm{E}+02,3.63 \mathrm{E}+02\),
    $2.65 \mathrm{E}+02,1.26 \mathrm{E}+02$,

Taper weighted geometric mean of the VPA populations:
$1.39 \mathrm{E}+04,1.16 \mathrm{E}+04,8.91 \mathrm{E}+03,5.41 \mathrm{E}+03,2.99 \mathrm{E}+03,1.48 \mathrm{E}+03,6.66 \mathrm{E}+02$, $2.70 \mathrm{E}+02,1.08 \mathrm{E}+02$,

Standard error of the weighted Log(VPA populations) :
.7107, .7233, . $7048, \quad .6634, \quad .6680, \quad .6456, \quad .6312, \quad .6644$,

Log catchability residuals.

Fleet : SUMMER SURVEY

| Age | , | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002, | 2003, | 2004, | 2005 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  | No dat | for | S fle | at t | s age |  |  |  |  |  |
| 2 | , | -.19, | .17, | . 32 , | -. 90, | .11, | . 64, | 1.10, | -. 05, | . 64, | . 52 |
| 3 | , | . 06 , | -. 30, | -. 68, | . 44, | -.49, | . 00 , | . 54, | -. 40, | . 01 , | . 38 |
| 4 |  | . 09 , | . 20 , | -. 72, | -. 24 , | -. 04 , | . 00 , | -.01, | . 02 , | -. 27 , | . 16 |
| 5 |  | . 59, | -. 15, | . 14, | -.79, | -.87, | -. 19, | . 04 , | -. 40, | . 38, | . 23 |
| 6 |  | . 08, | -. 26 , | . 49, | . 03, | -. 73, | -. 67, | -. 42, | -. 77, | . 21, | . 62 |
| 7 |  | . 21 , | -. 11, | -. 44 , | . 46 , | -. 03, | -. 40 , | -. 49, | -1.44, | . 07 , | . 44 |
| 8 |  | -. 20, | -. 34, | . 08 , | . 43, | -. 26 , | -. 10, | -. 55, | -1.13, | . 16, | . 49 |


| Age | , | 2006, | 2007, | 2008, | 2009, | 2010, | 2011, | 2012, | 2013, | 2014, | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  | No dat | for | s fle | at | s age |  |  |  |  |  |
| 2 | , | . 80, | -. 28 , | -1.80, | -.17, | . 65, | . 11, | -1.72, | -. 63, | . 27 , | . 44 |
| 3 | , | -. 10, | -. 68, | -. 58, | 1.05, | . 50 , | . 15 , | -. 97 , | -. 36 , | . 73 , | . 70 |
| 4 | , | -. 25, | -. 74 , | -. 88 , | . 43, | . 81, | . 48, | . 16, | -1.09, | . 26 , | 1.63 |
| 5 | , | -. 34, | -. 53, | -.11, | . 07 , | . 21 , | . 41, | . 32 , | . 38 , | -. 19, | . 82 |
| 6 |  | -. 42, | -. 43, | -.01, | . 51, | . 18, | . 14, | -. 20, | 1.00, | . 18, | . 50 |
| 7 |  | -.09, | -. 73, | -. 50, | . 44 , | . 38 , | . 28 , | -. 54, | . 62 , | . 06 , | . 39 |
| 8 |  | -.02, | -. 50, | -. 53, | . 26 , | . 01 , | -. 27 , | -. 36 , | . 31 , | -.01, | -. 05 |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age, | 2, | 3, | 4, | 5, | 6, | 7, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 8 Mean Log q, | -7.8746, | -6.7078, | -6.2936, | -6.0845, | -6.0486, | -6.0486, |
| -6.0486, | .7723, | .5560, | .6124, | .4451, | .4861, | .5222, |


#### Abstract

Ages with $q$ independent of year class strength and constant w.r.t. time.




Fleet : SPRING SURVEY (shift


| Age | , | 2006, | 2007, | 2008, | 2009, | 2010, | 2011, | 2012, | 2013, | 2014, | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  | -1.09, | . 12 , | . 43, | . 94, | . 36 , | -. 10, | -.03, | -. 35, | -.95, | . 00 |
| 2 | , | -. 66, | . 22 , | -.09, | . 76 , | . 54, | -. 20, | . 89, | -. 75, | -. 14, | 40 |
| 3 | , | -. 83, | . 10, | . 51, | . 33 , | . 47, | . 27 , | 1.13, | -. 53, | -. 48, | -. 35 |
| 4 | , | -. 79, | -.03, | . 66 , | . 35 , | . 58, | -. 36, | 1.12, | -. 17, | -. 08, | . 45 |
| 5 |  | -. 36 , | -. 16 , | . 47 , | . 25 , | . 61 , | -. 65, | . 36 , | -. 07, | . 19, | -. 36 |
| 6 |  | -. 39, | -. 05 , | . 04 , | -.03, | 1.12, | -. 79 , | -. 29 , | -. 19, | -. 06 , | -. 20 |
| 7 |  | -. 30, | -. 50, | -. 10, | . 09 , | . 72 , | -. 33, | -.13, | -.16, | -. 49, | -. 94 |
| 8 |  | 30, | -. 46 , | 42, | 11, | 24 | 1.25, | 51, | -. 08, | -. 79 , | 1.4 |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age, | 1, | 2, | 3, | 4, | 5, | 6, |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean $Q$

| 1, | 1.00, | -.024, | 8.33, | .57, | 23, | .72, | -8.34, |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2, | 1.04, | -.221, | 6.81, | .65, | 23, | .60, | -6.89, |
| 3, | .89, | .833, | 6.29, | .75, | 23, | .44, | -6.00, |
| 4, | .91, | .756, | 5.95, | .77, | 23, | .42, | -5.71, |
| 5, | .90, | .858, | 5.93, | .79, | 23, | .38, | -5.74, |
| 6, | .89, | .776, | 6.10, | .71, | 23, | .41, | -5.98, |


| 7, | .97, | .199, | 6.18, | .75, | 23, | .39, | -6.18, |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 8, | .64, | 1.579, | 6.03, | .48, | 23, | .72, | -6.47, |

Terminal year survivor and $F$ summaries :
Age 1 Catchability constant w.r.t. time and dependent on age
Year class = 2014

| Fleet, Estimated | Estimated, | Int, | Ext, | Var, |  | Scaled, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| , | Survivors, | s.e, | s.e, | Ratio, | , | Weights |
| F |  |  |  |  |  |  |
| $\begin{aligned} & \text { SUMMER SURVEY , } \\ & .000 \end{aligned}$ | 1., | . 000 , | . 000, | . 00 , | 0 , | . 000 , |
| $\begin{aligned} & \text { SPRING SURVEY (shift, } \\ & .000 \end{aligned}$ | 3374., | . 717, | . 000 , | . 00 , | 1, | 1.000, |
| F shrinkage mean , $.000$ | 0., | 2.00, , , |  |  |  | . 000, |
| Weighted prediction : |  |  |  |  |  |  |
| Survivors, Int, at end of year, s.e, 3374., .72, | Ext, s.e, .00, | $\begin{array}{ll} \text { N, } & \text { Var, } \\ \text { ', } & \text { Ratio, } \\ \text { 1, } & .000, \end{array}$ | F $.000$ |  |  |  |

Age 2 Catchability constant w.r.t. time and dependent on age

Year class $=2013$

| Fleet, Estimated | Estimated, | Int, | Ext, | Var, | N, | Scaled, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| , | Survivors, | S.e, | S.e, | Ratio, | , | Weights, |
| F |  |  |  |  |  |  |
| SUMMER SURVEY $.049$ | 2810., | . 791 , | . 000 , | . 00 , | 1, | . 236 , |
| $\begin{aligned} & \text { SPRING SURVEY (shift, } \\ & .085 \end{aligned}$ | 1586., | . 451, | . 659 , | 1.46, | 2, | . 725 , |
| F shrinkage mean , | 1576., | 2.00, |  |  |  | . 040 , |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $1815 .$, | .38, | .35, | 4, | .917, | .075 |

Age 3 Catchability constant w.r.t. time and dependent on age
Year class $=2012$



Age 5 Catchability constant w.r.t. time and dependent on age
Year class $=2010$

| Fleet, Estimated | Estimated, | Int, | Ext, | Var, | N, | Scaled, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| , | Survivors, | s.e, | s.e, | Ratio, | , | Weights, |
| F |  |  |  |  |  |  |
| $\begin{aligned} & \text { SUMMER SURVEY } \\ & .393 \end{aligned}$ | 394. | . 296 , | . 459, | 1.55, | 4, | . 395, |
| $\begin{aligned} & \text { SPRING SURVEY (shift, } \\ & .529 \end{aligned}$ | 273., | . 236, | . 212, | . 90 , | 5, | . 588, |
| F shrinkage mean | 267., | 2.00, |  |  |  | . 017 , |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $315 .$, | .18, | .21, | $10^{\prime}$, | 1.127, | .472 |

Age 6 Catchability constant w.r.t. time and dependent on age

Year class $=2009$


Age 7 Catchability constant w.r.t. time and age (fixed at the value for age) 6


Table 4.6.2. Faroe Plateau cod (subdivision 5.b.1). Fishing mortality-at-age from the XSA model.

| Year \age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | FBAR 3-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1959 | 0.1829 | 0.4853 | 0.4463 | 0.6303 | 0.3909 | 0.6060 | 0.3005 | 0.4784 | 0.4784 | 0.5117 |
| 1960 | 0.4570 | 0.6793 | 0.6222 | 0.5290 | 0.7826 | 0.6920 | 1.0328 | 0.7389 | 0.7389 | 0.6610 |
| 1961 | 0.3346 | 0.5141 | 0.4986 | 0.5737 | 0.4863 | 0.9566 | 0.8116 | 0.6715 | 0.6715 | 0.6059 |
| 1962 | 0.2701 | 0.4982 | 0.4838 | 0.7076 | 0.5569 | 0.3662 | 0.6826 | 0.5641 | 0.5641 | 0.5226 |
| 1963 | 0.2534 | 0.4138 | 0.5172 | 0.5124 | 0.5405 | 0.4879 | 0.3269 | 0.4806 | 0.4806 | 0.4944 |
| 1964 | 0.1086 | 0.2997 | 0.4523 | 0.5229 | 0.5659 | 0.6677 | 0.3531 | 0.5164 | 0.5164 | 0.5017 |
| 1965 | 0.1209 | 0.2518 | 0.4498 | 0.5622 | 0.6604 | 0.5305 | 0.4345 | 0.5318 | 0.5318 | 0.4909 |
| 1966 | 0.0829 | 0.1969 | 0.2552 | 0.4499 | 0.5016 | 0.9680 | 0.8520 | 0.6106 | 0.6106 | 0.4743 |
| 1967 | 0.0789 | 0.2389 | 0.2687 | 0.3442 | 0.5779 | 0.5203 | 1.0438 | 0.5556 | 0.5556 | 0.3900 |
| 1968 | 0.1010 | 0.2318 | 0.3949 | 0.5339 | 0.4472 | 0.7132 | 0.3331 | 0.4882 | 0.4882 | 0.4642 |
| 1969 | 0.1099 | 0.3063 | 0.3806 | 0.4180 | 0.5709 | 0.5118 | 0.8457 | 0.5499 | 0.5499 | 0.4375 |
| 1970 | 0.0530 | 0.2081 | 0.3654 | 0.3409 | 0.3709 | 0.6559 | 0.4208 | 0.4339 | 0.4339 | 0.3882 |
| 1971 | 0.0309 | 0.1337 | 0.2225 | 0.3845 | 0.5572 | 0.4651 | 0.7528 | 0.4800 | 0.4800 | 0.3526 |
| 1972 | 0.0464 | 0.1476 | 0.2070 | 0.2497 | 0.6058 | 0.4686 | 0.2464 | 0.3578 | 0.3578 | 0.3358 |
| 1973 | 0.0657 | 0.2322 | 0.3048 | 0.2813 | 0.2526 | 0.3722 | 0.3259 | 0.3091 | 0.3091 | 0.2886 |
| 1974 | 0.081 | 0.1568 | 0.2046 | 0.2953 | 0.3797 | 0.5330 | 0.3052 | 0.3457 | 0.3457 | 0.3139 |
| 1975 | 0.0774 | 0.3193 | 0.4359 | 0.4134 | 0.4544 | 0.3504 | 0.4485 | 0.4235 | 0.4235 | 0.3947 |
| 1976 | 0.0933 | 0.1723 | 0.3665 | 0.5568 | 0.5167 | 0.7619 | 0.6429 | 0.5738 | 0.5738 | 0.4749 |
| 1977 | 0.0481 | 0.3036 | 0.4748 | 0.7532 | 0.7333 | 1.1138 | 0.7776 | 0.7783 | 0.7783 | 0.6757 |
| 1978 | 0.0588 | 0.1896 | 0.4291 | 0.4289 | 0.4850 | 0.5968 | 0.5674 | 0.5054 | 0.5054 | 0.4259 |
| 1979 | 0.0433 | 0.2623 | 0.4309 | 0.5049 | 0.4906 | 0.4480 | 0.6903 | 0.5170 | 0.5170 | 0.4273 |
| 1980 | 0.054 | 0.2391 | 0.3695 | 0.4337 | 0.5182 | 0.4119 | 0.6437 | 0.4790 | 0.4790 | 0.3945 |
| 1981 | 0.0523 | 0.2877 | 0.3409 | 0.4369 | 0.5644 | 0.6940 | 0.5015 | 0.5115 | 0.5115 | 0.4648 |
| 1982 | 0.0586 | 0.2227 | 0.3602 | 0.3887 | 0.4047 | 0.6926 | 0.5526 | 0.4834 | 0.4834 | 0.4138 |
| 1983 | 0.0991 | 0.4672 | 0.5585 | 0.6411 | 0.7835 | 1.0779 | 0.9417 | 0.8087 | 0.8087 | 0.7056 |
| 1984 | 0.1073 | 0.3712 | 0.5790 | 0.6609 | 0.4533 | 0.4761 | 0.4792 | 0.5340 | 0.5340 | 0.5081 |
| 1985 | 0.0658 | 0.3543 | 0.5075 | 0.6134 | 0.9235 | 1.1081 | 1.3203 | 0.9042 | 0.9042 | 0.7014 |
| 1986 | 0.0247 | 0.3545 | 0.6225 | 0.7030 | 0.8256 | 0.8400 | 0.5408 | 0.7131 | 0.7131 | 0.6691 |
| 1987 | 0.0291 | 0.2208 | 0.4754 | 0.4850 | 0.5555 | 0.4896 | 0.6222 | 0.5298 | 0.5298 | 0.4453 |
| 1988 | 0.0666 | 0.3531 | 0.5639 | 0.5490 | 0.7735 | 0.7980 | 0.8641 | 0.7165 | 0.7165 | 0.6075 |
| 1989 | 0.1659 | 0.4394 | 0.7614 | 0.7620 | 0.9617 | 1.0574 | 1.0994 | 0.9386 | 0.9386 | 0.7964 |
| 1990 | 0.0780 | 0.3355 | 0.6371 | 0.8016 | 0.7141 | 0.8518 | 1.1365 | 0.8369 | 0.8369 | 0.6680 |
| 1991 | 0.0324 | 0.1997 | 0.4499 | 0.5980 | 0.7462 | 0.5816 | 0.7180 | 0.6242 | 0.6242 | 0.5151 |
| 1992 | 0.0201 | 0.1001 | 0.3270 | 0.3478 | 0.6366 | 0.8918 | 0.4457 | 0.5342 | 0.5342 | 0.4607 |
| 1993 | 0.0132 | 0.1021 | 0.1869 | 0.2550 | 0.2027 | 0.4403 | 0.5790 | 0.3350 | 0.3350 | 0.2374 |
| 1994 | 0.0254 | 0.1130 | 0.1910 | 0.2503 | 0.2228 | 0.1588 | 0.3209 | 1.1015 | 1.1015 | 0.1872 |
| 1995 | 0.0698 | 0.1615 | 0.4658 | 0.2809 | 0.3620 | 0.3394 | 0.2344 | 0.7368 | 0.7368 | 0.3219 |
| 1996 | 0.0301 | 0.1916 | 0.4512 | 0.8130 | 0.9099 | 1.1481 | 0.9542 | 1.0211 | 1.0211 | 0.7028 |
| 1997 | 0.0341 | 0.1459 | 0.4082 | 0.8299 | 1.0584 | 1.4231 | 1.3855 | 1.1287 | 1.1287 | 0.7731 |
| 1998 | 0.0866 | 0.1719 | 0.2665 | 0.6381 | 1.0420 | 0.8124 | 1.2396 | 0.9235 | 0.9235 | 0.5862 |
| 1999 | 0.0989 | 0.2922 | 0.2987 | 0.3267 | 0.6857 | 1.1262 | 0.9085 | 0.6112 | 0.6112 | 0.5459 |


| YEAR $\backslash$ AGE | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ | FBAR 3- 7 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 0.1279 | 0.3268 | 0.3887 | 0.2539 | 0.3338 | 0.5659 | 0.9070 | 0.2657 | 0.2657 | 0.3738 |
| 2001 | 0.1586 | 0.3465 | 0.4580 | 0.3104 | 0.3517 | 0.7000 | 0.6738 | 0.9194 | 0.9194 | 0.4333 |
| 2002 | 0.1939 | 0.4930 | 0.6010 | 0.8261 | 0.8322 | 1.3642 | 1.2352 | 1.4999 | 1.4999 | 0.8233 |
| 2003 | 0.1347 | 0.3177 | 0.6873 | 0.8792 | 0.9433 | 0.9388 | 0.9717 | 1.8512 | 1.8512 | 0.7532 |
| 2004 | 0.0312 | 0.1864 | 0.2971 | 0.7491 | 0.9767 | 1.1465 | 1.0737 | 2.0794 | 2.0794 | 0.6712 |
| 2005 | 0.0995 | 0.2725 | 0.4053 | 0.5004 | 0.8175 | 0.8836 | 0.6153 | 1.2362 | 1.2362 | 0.5758 |
| 2006 | 0.1815 | 0.3260 | 0.3524 | 0.6008 | 0.8083 | 0.9459 | 1.0038 | 0.2921 | 0.2921 | 0.6067 |
| 2007 | 0.1179 | 0.3125 | 0.3714 | 0.4256 | 0.5770 | 0.6611 | 0.6755 | 0.8110 | 0.8110 | 0.4695 |
| 2008 | 0.0503 | 0.2611 | 0.3371 | 0.3956 | 0.4856 | 0.7299 | 1.2043 | 1.5623 | 1.5623 | 0.4419 |
| 2009 | 0.1246 | 0.6865 | 0.5315 | 0.4886 | 0.5243 | 0.4263 | 0.6160 | 1.5416 | 1.5416 | 0.5314 |
| 2010 | 0.2314 | 0.4741 | 0.6140 | 0.6268 | 0.9135 | 0.7058 | 0.5852 | 0.9782 | 0.9782 | 0.6668 |
| 2011 | 0.0882 | 0.4349 | 0.5969 | 0.5779 | 0.5580 | 0.6496 | 0.4336 | 0.3831 | 0.3831 | 0.5635 |
| 2012 | 0.0292 | 0.1954 | 0.4817 | 0.5952 | 0.7415 | 0.6551 | 0.9463 | 0.4027 | 0.4027 | 0.5338 |
| 2013 | 0.0213 | 0.1365 | 0.1861 | 0.3999 | 0.3758 | 0.3035 | 0.3360 | 0.5422 | 0.5422 | 0.2804 |
| 2014 | 0.0575 | 0.2839 | 0.3501 | 0.4690 | 0.4433 | 0.3199 | 0.1690 | 0.1931 | 0.1931 | 0.3732 |
| 2015 | 0.0749 | 0.2148 | 0.5177 | 0.4715 | 0.5328 | 0.5624 | 0.2899 | 0.4610 | 0.4610 | 0.4599 |

Table 4.6.3. Faroe Plateau cod (subdivision 5.b.1). Stock number-at-age from the XSA model.

| Year \AGe | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1959 | 17399 | 13238 | 12185 | 2634 | 4092 | 683 | 503 | 213 | 29 | 0 | 50976 |
| 1960 | 14680 | 14245 | 9027 | 6141 | 1380 | 1784 | 378 | 225 | 129 | 0 | 47989 |
| 1961 | 25227 | 12019 | 7385 | 3747 | 2699 | 666 | 668 | 155 | 66 | 0 | 52630 |
| 1962 | 24782 | 20654 | 7042 | 3616 | 1863 | 1245 | 335 | 210 | 56 | 0 | 59804 |
| 1963 | 26668 | 20290 | 12907 | 3503 | 1825 | 752 | 584 | 190 | 87 | 0 | 66807 |
| 1964 | 10100 | 21834 | 12893 | 6986 | 1710 | 895 | 358 | 294 | 112 | 0 | 55183 |
| 1965 | 22676 | 8269 | 16037 | 7823 | 3639 | 830 | 416 | 151 | 169 | 0 | 60009 |
| 1966 | 28643 | 18566 | 5999 | 10207 | 4085 | 1698 | 351 | 200 | 80 | 0 | 69829 |
| 1967 | 21475 | 23451 | 13990 | 4034 | 6475 | 2133 | 842 | 109 | 70 | 0 | 72579 |
| 1968 | 11390 | 17582 | 17744 | 9020 | 2525 | 3757 | 980 | 410 | 31 | 0 | 63439 |
| 1969 | 10514 | 9325 | 13012 | 11522 | 4976 | 1212 | 1967 | 393 | 240 | 0 | 53161 |
| 1970 | 14569 | 8608 | 6840 | 7843 | 6447 | 2682 | 561 | 965 | 138 | 0 | 48654 |
| 1971 | 26041 | 11928 | 6684 | 4548 | 4456 | 3754 | 1516 | 238 | 519 | 0 | 59683 |
| 1972 | 15356 | 21320 | 9469 | 4788 | 2981 | 2483 | 1760 | 779 | 92 | 0 | 59029 |
| 1973 | 37229 | 12573 | 16664 | 6689 | 3187 | 1901 | 1109 | 902 | 499 | 400 | 81153 |
| 1974 | 46803 | 30480 | 9639 | 10816 | 4037 | 1969 | 1209 | 626 | 533 | 342 | 106456 |
| 1975 | 22687 | 38319 | 23000 | 6747 | 7217 | 2460 | 1103 | 581 | 378 | 476 | 102968 |
| 1976 | 12208 | 18575 | 29035 | 13683 | 3572 | 3908 | 1279 | 636 | 304 | 466 | 83665 |
| 1977 | 13128 | 9995 | 13853 | 20010 | 7765 | 1676 | 1909 | 489 | 274 | 18 | 69116 |
| 1978 | 18318 | 10748 | 7799 | 8372 | 10190 | 2993 | 659 | 513 | 184 | 154 | 59931 |
| 1979 | 28804 | 14998 | 8298 | 5282 | 4463 | 5433 | 1509 | 297 | 238 | 103 | 69424 |
| 1980 | 17100 | 23582 | 11759 | 5226 | 2811 | 2206 | 2723 | 789 | 122 | 52 | 66370 |
| 1981 | 27027 | 14000 | 18286 | 7580 | 2957 | 1491 | 1076 | 1477 | 339 | 150 | 74384 |
| 1982 | 30732 | 22128 | 10878 | 11228 | 4413 | 1564 | 694 | 440 | 732 | 348 | 83159 |
| 1983 | 58342 | 25161 | 17086 | 7128 | 6412 | 2450 | 854 | 284 | 207 | 200 | 118126 |
| 1984 | 21157 | 47766 | 18656 | 8767 | 3339 | 2765 | 916 | 238 | 91 | 174 | 103870 |
| 1985 | 11616 | 17322 | 35130 | 10538 | 4023 | 1412 | 1439 | 466 | 121 | 146 | 82212 |
| 1986 | 12108 | 9511 | 13279 | 20181 | 5194 | 1784 | 459 | 389 | 102 | 81 | 63087 |
| 1987 | 10661 | 9913 | 7597 | 7627 | 8866 | 2105 | 640 | 162 | 185 | 69 | 47826 |
| 1988 | 19749 | 8729 | 7884 | 4987 | 3882 | 4469 | 989 | 321 | 71 | 53 | 51135 |
| 1989 | 4441 | 16169 | 6686 | 4535 | 2323 | 1835 | 1688 | 365 | 111 | 16 | 38170 |
| 1990 | 8132 | 3636 | 11214 | 3528 | 1734 | 888 | 574 | 480 | 99 | 50 | 30336 |
| 1991 | 13900 | 6658 | 2754 | 6565 | 1527 | 637 | 356 | 201 | 126 | 56 | 32780 |
| 1992 | 12320 | 11380 | 5277 | 1846 | 3428 | 688 | 247 | 163 | 80 | 90 | 35520 |
| 1993 | 30804 | 10087 | 9132 | 3909 | 1090 | 1982 | 298 | 83 | 85 | 96 | 57567 |
| 1994 | 52357 | 25220 | 8150 | 6751 | 2655 | 692 | 1325 | 157 | 38 | 26 | 97371 |
| 1995 | 15991 | 42867 | 20130 | 5960 | 4566 | 1692 | 453 | 926 | 93 | 102 | 92780 |
| 1996 | 8049 | 13092 | 32730 | 14024 | 3062 | 2823 | 965 | 264 | 599 | 80 | 75688 |
| 1997 | 7407 | 6590 | 10401 | 22125 | 7312 | 1112 | 930 | 251 | 83 | 192 | 56403 |
| 1998 | 17868 | 6064 | 5214 | 7359 | 12043 | 2611 | 316 | 184 | 51 | 45 | 51755 |
| 1999 | 24393 | 14629 | 4553 | 3595 | 4616 | 5209 | 754 | 115 | 44 | 19 | 57926 |


| YEAR\AGE | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ | TOTAL |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2000 | 36505 | 19972 | 10850 | 2783 | 2183 | 2726 | 2148 | 200 | 38 | 5 | 77409 |
| 2001 | 16074 | 29888 | 14388 | 6406 | 1545 | 1387 | 1598 | 999 | 66 | 11 | 72361 |
| 2002 | 7466 | 13160 | 20880 | 8331 | 3318 | 927 | 799 | 650 | 417 | 10 | 55957 |
| 2003 | 4308 | 6112 | 8875 | 10441 | 3739 | 1189 | 330 | 167 | 155 | 27 | 35344 |
| 2004 | 7182 | 3527 | 4374 | 5288 | 4300 | 1271 | 379 | 106 | 52 | 44 | 26522 |
| 2005 | 9029 | 5880 | 2799 | 2972 | 3217 | 1664 | 392 | 99 | 30 | 47 | 26129 |
| 2006 | 6143 | 7393 | 4358 | 1745 | 1623 | 1597 | 602 | 133 | 44 | 13 | 23649 |
| 2007 | 7896 | 5030 | 5048 | 2576 | 1004 | 728 | 583 | 191 | 40 | 6 | 23102 |
| 2008 | 10060 | 6465 | 3660 | 3024 | 1454 | 537 | 335 | 246 | 80 | 26 | 25887 |
| 2009 | 13806 | 8237 | 5033 | 2308 | 1767 | 802 | 271 | 132 | 60 | 23 | 32439 |
| 2010 | 5244 | 11303 | 5954 | 2074 | 1111 | 888 | 389 | 145 | 58 | 66 | 27231 |
| 2011 | 2301 | 4294 | 7343 | 3034 | 919 | 486 | 291 | 157 | 66 | 7 | 18898 |
| 2012 | 3530 | 1884 | 3218 | 3892 | 1368 | 422 | 228 | 125 | 83 | 73 | 14822 |
| 2013 | 9349 | 2890 | 1498 | 2167 | 1968 | 617 | 165 | 97 | 40 | 13 | 18804 |
| 2014 | 2918 | 7654 | 2316 | 1070 | 1473 | 1080 | 347 | 100 | 57 | 6 | 17022 |
| 2015 | 4121 | 2389 | 5917 | 1428 | 617 | 755 | 568 | 206 | 69 | 21 | 16090 |
| 2016 | 0 | 3374 | 1815 | 3908 | 697 | 315 | 363 | 265 | 126 | 46 | 10908 |

Table 4.6.4. Faroe Plateau cod (subdivision 5.b.1). Summary table from the XSA model. The re-sults from the short-term prediction are shown in bold.

| Year | Recruitment | BIomass | Bıomass | BIomass | LANDINGS | Mean F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 2 | Age $2+$ | Age 3+ | SSB |  | Ages 3-7 |
|  | THOUSANDS | TONNES | TONNES | TONNES | TONNES |  |
| 1959 | 13238 | 67803 | 56550 | 48869 | 22415 | 0.5117 |
| 1960 | 14245 | 75862 | 61619 | 54447 | 32255 | 0.661 |
| 1961 | 12019 | 65428 | 52459 | 46439 | 21598 | 0.6059 |
| 1962 | 20654 | 68225 | 47568 | 43326 | 20967 | 0.5226 |
| 1963 | 20290 | 77602 | 56500 | 49054 | 22215 | 0.4944 |
| 1964 | 21834 | 84666 | 63483 | 55362 | 21078 | 0.5017 |
| 1965 | 8269 | 75043 | 67442 | 57057 | 24212 | 0.4909 |
| 1966 | 18566 | 83919 | 65724 | 60629 | 20418 | 0.4743 |
| 1967 | 23451 | 105289 | 82778 | 73934 | 23562 | 0.39 |
| 1968 | 17582 | 110433 | 94958 | 82484 | 29930 | 0.4642 |
| 1969 | 9325 | 105537 | 95372 | 83487 | 32371 | 0.4375 |
| 1970 | 8608 | 98398 | 90131 | 82035 | 24183 | 0.3882 |
| 1971 | 11928 | 78218 | 68559 | 63308 | 23010 | 0.3526 |
| 1972 | 21320 | 76439 | 62363 | 57180 | 18727 | 0.3358 |
| 1973 | 12573 | 110713 | 96756 | 83547 | 22228 | 0.2886 |
| 1974 | 30480 | 139266 | 106341 | 98434 | 24581 | 0.3139 |
| 1975 | 38319 | 153664 | 123391 | 109566 | 36775 | 0.3947 |
| 1976 | 18575 | 161260 | 143807 | 123077 | 39799 | 0.4749 |
| 1977 | 9995 | 136211 | 127520 | 112057 | 34927 | 0.6757 |
| 1978 | 10748 | 96227 | 84269 | 78497 | 26585 | 0.4259 |
| 1979 | 14998 | 85112 | 71659 | 66723 | 23112 | 0.4273 |
| 1980 | 23582 | 85038 | 63178 | 58887 | 20513 | 0.3945 |
| 1981 | 14000 | 88411 | 73287 | 63562 | 22963 | 0.4648 |
| 1982 | 22128 | 98963 | 71739 | 67033 | 21489 | 0.4138 |
| 1983 | 25161 | 123255 | 89581 | 78542 | 38133 | 0.7056 |
| 1984 | 47766 | 152158 | 95072 | 96773 | 36979 | 0.5081 |
| 1985 | 17322 | 131240 | 115566 | 84786 | 39484 | 0.7014 |
| 1986 | 9511 | 99271 | 88821 | 73693 | 34595 | 0.6691 |
| 1987 | 9913 | 78362 | 67522 | 62241 | 21391 | 0.4453 |
| 1988 | 8729 | 66177 | 56912 | 52125 | 23182 | 0.6075 |
| 1989 | 16169 | 59031 | 42701 | 38406 | 22068 | 0.7964 |
| 1990 | 3636 | 38276 | 34837 | 29270 | 13692 | 0.668 |
| 1991 | 6658 | 28679 | 23491 | 21069 | 8750 | 0.5151 |
| 1992 | 11380 | 35684 | 24430 | 20755 | 6396 | 0.4607 |
| 1993 | 10087 | 51034 | 39378 | 33068 | 6107 | 0.2374 |
| 1994 | 25220 | 83914 | 53804 | 42475 | 9046 | 0.1872 |
| 1995 | 42867 | 144645 | 92428 | 54320 | 23045 | 0.3219 |
| 1996 | 13092 | 143005 | 129704 | 85321 | 40422 | 0.7028 |
| 1997 | 6590 | 97233 | 91293 | 81714 | 34304 | 0.7731 |
| 1998 | 6064 | 66872 | 60784 | 56284 | 24005 | 0.5862 |
| 1999 | 14629 | 66269 | 50916 | 45830 | 19245 | 0.5459 |


| Year | Recruitment | Biomass | Biomass | Biomass | LANDINGS | Mean F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 2 | Age $2+$ | Age 3+ | SSB |  | Ages 3-7 |
|  | THousands | TONNES | TONNES | TONNES | TONNES |  |
| 2000 | 19972 | 91995 | 63718 | 46396 | 21833 | 0.3738 |
| 2001 | 29888 | 110410 | 75622 | 59118 | 28577 | 0.4333 |
| 2002 | 13160 | 98445 | 85066 | 56006 | 38834 | 0.8233 |
| 2003 | 6112 | 60392 | 55378 | 40542 | 25167 | 0.7532 |
| 2004 | 3527 | 36140 | 32490 | 26435 | 12840 | 0.6712 |
| 2005 | 5880 | 31066 | 25273 | 22942 | 10119 | 0.5758 |
| 2006 | 7393 | 28949 | 22757 | 19879 | 9844 | 0.6067 |
| 2007 | 5030 | 26543 | 21832 | 16786 | 7511 | 0.4695 |
| 2008 | 6465 | 29423 | 21607 | 20129 | 7315 | 0.4419 |
| 2009 | 8237 | 29617 | 22981 | 19359 | 9979 | 0.5314 |
| 2010 | 11303 | 37225 | 25370 | 21047 | 12757 | 0.6668 |
| 2011 | 4294 | 29310 | 25810 | 18135 | 9692 | 0.5635 |
| 2012 | 1884 | 22838 | 20946 | 17848 | 7204 | 0.5338 |
| 2013 | 2890 | 22114 | 19196 | 19083 | 4473 | 0.2804 |
| 2014 | 7654 | 27567 | 19156 | 20087 | 5715 | 0.3732 |
| 2015 | 2389 | 28520 | 25663 | 19729 | 7394 | 0.4599 |
| 2016 | 3374 | 26625 | 25663 | 22408 | 7514 | 0.4599 |
| 2017 | 4311 | 25745 |  | 20162 | 7170 | 0.4599 |
| 2018 | 4311 | 25020 |  | 18911 |  |  |
| Average | 14736 | 79908 | 64677 | 55347 | 21832 | 0.5102 |

Table 4.6.5. Faroe Plateau cod (subdivision 5.b.1). Biomass (age $2+$, tons) from 1710-1859 based on two approaches. The left part of the table is modelled by scaled dried cod export (taking account of increasing tendency to export more fish over time) 1710-1859. The right part of the table is modelled by the same model, but fitted for the 1710-1840 period, and a separate model, a factor, is used for the 1841-1859 period. This was an attempt to take account of a possible increased dried cod export caused by better possibilities of fishers to sell their landings. The value in 1860 (taken from next table) is also shown. Missing years are modelled by linear interpolation. Year label = first row + first column.

|  | 1700 | 1750 | 1800 | 1850 | 1700 | 1750 | 1800 | 1850 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  | 203118 | 145689 | 163103 |  | 209334 | 153202 | 69133 |
| 1 |  | 150800 | 150618 | 168040 |  | 154170 | 158624 | 75653 |
| 2 |  | 100949 | 109427 | 175508 |  | 100590 | 113557 | 86379 |
| 3 |  | 74141 | 81679 | 181495 |  | 72481 | 83220 | 95999 |
| 4 |  | 98616 | 70210 | 205909 |  | 97678 | 70898 | 142806 |
| 5 |  | 62191 | 89050 | 236862 |  | 60147 | 91333 | 215365 |
| 6 |  | 131926 | 100897 | 210540 |  | 134388 | 104274 | 193343 |
| 7 |  | 141252 | 115479 | 184218 |  | 144056 | 120134 | 171320 |
| 8 |  | 180782 | 113116 | 157896 |  | 186010 | 117581 | 149298 |
| 9 |  | 159904 | 93431 | 131574 |  | 163392 | 96800 | 127275 |
| 10 | 142041 | 133776 | 73745 | 105252 | 139327 | 135636 | 76018 | 105252 |
| 11 | 168323 | 107648 | 54060 |  | 166755 | 107881 | 55236 |  |
| 12 | 193811 | 120502 | 34374 |  | 193421 | 121629 | 34455 |  |
| 13 | 193026 | 120502 | 14689 |  | 192735 | 121629 | 13673 |  |
| 14 | 154233 | 128892 | 67502 |  | 152771 | 130692 | 69965 |  |
| 15 | 109779 | 145077 | 90579 |  | 107070 | 148357 | 94857 |  |
| 16 | 58303 | 139000 | 111302 |  | 54676 | 141834 | 116549 |  |
| 17 | 70962 | 129309 | 113210 |  | 67616 | 131610 | 118767 |  |
| 18 | 98712 | 111100 | 130165 |  | 96031 | 112061 | 138050 |  |
| 19 | 127083 | 95355 | 173807 |  | 124525 | 95335 | 187049 |  |
| 20 | 114273 | 103794 | 183183 |  | 111543 | 104507 | 197681 |  |
| 21 | 98139 | 117533 | 151025 |  | 94963 | 119275 | 161727 |  |
| 22 | 77036 | 117533 | 128521 |  | 73459 | 119275 | 136272 |  |
| 23 | 65720 | 131271 | 121350 |  | 62238 | 134044 | 128310 |  |
| 24 | 54404 | 94983 | 119010 |  | 51018 | 95739 | 125814 |  |
| 25 | 54404 | 58694 | 113337 |  | 51018 | 57434 | 119552 |  |
| 26 | 55649 | 71518 | 102938 |  | 52294 | 70870 | 108042 |  |
| 27 | 86296 | 79834 | 116320 |  | 83531 | 79659 | 122965 |  |
| 28 | 101645 | 90404 | 124944 |  | 99323 | 90772 | 132793 |  |
| 29 | 120159 | 96467 | 136327 |  | 118226 | 97236 | 145626 |  |
| 30 | 122389 | 104086 | 150109 |  | 120593 | 105556 | 161352 |  |
| 31 | 112436 | 111704 | 161142 |  | 110278 | 113876 | 174088 |  |
| 32 | 86199 | 92367 | 169340 |  | 83857 | 93305 | 183519 |  |
| 33 | 59963 | 92367 | 171628 |  | 57435 | 93305 | 186276 |  |
| 34 | 33726 | 85472 | 157524 |  | 31014 | 86016 | 170288 |  |
| 35 | 49250 | 102178 | 153222 |  | 46434 | 103919 | 165504 |  |
| 36 | 81195 | 100998 | 154433 |  | 79333 | 102688 | 167039 |  |
| 37 | 86494 | 120264 | 159721 |  | 84577 | 123750 | 173188 |  |


|  | 1700 | 1750 | 1800 | 1850 | 1700 | 1750 | 1800 | 1850 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 38 | 79128 | 146668 | 141223 | 77256 | 152842 | 152440 |  |  |
| 39 | 61034 | 172556 | 135717 | 58429 | 181282 | 146245 |  |  |
| 40 | 50622 | 164087 | 156283 | 48317 | 172091 | 135996 |  |  |
| 41 | 54594 | 138988 | 219704 | 52530 | 144576 | 199136 |  |  |
| 42 | 18387 | 122896 | 238434 | 16411 | 126933 | 187345 |  |  |
| 43 | 103720 | 117417 | 243695 | 102334 | 121048 | 199009 |  |  |
| 44 | 83076 | 124089 | 223875 | 81115 | 128422 | 154909 |  |  |
| 45 | 83076 | 130899 | 210972 | 81115 | 136024 | 136441 |  |  |
| 46 | 88182 | 113963 | 188260 | 86580 | 117918 | 101729 |  |  |
| 47 | 186414 | 96159 | 184208 | 191549 | 98703 | 95822 |  |  |
| 48 | 193149 | 77125 | 177024 | 198624 | 77981 | 88167 |  |  |
| 49 | 237895 | 117737 | 175956 | 247009 | 122529 | 87678 |  |  |

Table 4.6.6. Faroe Plateau cod (subdivision 5.b.1). Biomass (age 2+, tons) from 1860-2015. The biomass from 1860-1905 is based on scaled cpue from Faroese and Shetland vessels. The biomass from 1906-1958 is based on scaled cpue from British steam trawlers. The results from the age-based assessment from 1959-2015 are shown for completeness. Year label = first row + first column.

|  | 1850 | 1875 | 1900 | 1925 | 1950 | 1975 | 2000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  | 236420 | 143952 | 129353 | 152207 | 153664 | 91995 |
| 1 |  | 224948 | 156472 | 185574 | 124325 | 161260 | 110410 |
| 2 |  | 208758 | 151509 | 162034 | 116783 | 136211 | 98445 |
| 3 |  | 213729 | 163172 | 126611 | 116783 | 96227 | 60392 |
| 4 |  | 199256 | 145938 | 135524 | 146493 | 85112 | 36140 |
| 5 |  | 241850 | 130638 | 142608 | 149464 | 85038 | 31066 |
| 6 |  | 254217 | 125162 | 139409 | 108327 | 88411 | 28949 |
| 7 |  | 257567 | 148793 | 121354 | 112898 | 98963 | 26543 |
| 8 |  | 221818 | 108532 | 108327 | 84102 | 123255 | 29423 |
| 9 |  | 173547 | 175051 | 107870 | 67803 | 152158 | 29617 |
| 10 | 105252 | 102923 | 149669 | 91187 | 75862 | 131240 | 37225 |
| 11 | 88276 | 86780 | 175051 | 102385 | 65428 | 99271 | 29310 |
| 12 | 95349 | 72702 | 161922 | 95758 | 68225 | 78362 | 22838 |
| 13 | 127038 | 71090 | 137415 | 93244 | 77602 | 66177 | 22114 |
| 14 | 152220 | 84268 | 112908 | 143439 | 84666 | 59031 | 27567 |
| 15 | 118946 | 166257 | 122391 | 193635 | 75043 | 38276 | 28520 |
| 16 | 107629 | 253036 | 143731 | 216611 | 83919 | 28679 |  |
| 17 | 94463 | 258572 | 171332 | 188465 | 105289 | 35684 |  |
| 18 | 149274 | 187962 | 213691 | 196270 | 110433 | 51034 |  |
| 19 | 198563 | 137071 | 205685 | 210683 | 105537 | 83914 |  |
| 20 | 181645 | 195446 | 98904 | 225096 | 98398 | 144645 |  |
| 21 | 139371 | 184553 | 117284 | 239509 | 78218 | 143005 |  |
| 22 | 108056 | 155015 | 160172 | 177346 | 76439 | 97233 |  |
| 23 | 178060 | 104436 | 133039 | 122497 | 110713 | 66872 |  |
| 24 | 232567 | 133637 | 136895 | 164777 | 139266 | 66269 |  |

Table 4.7.1. Faroe Plateau cod (subdivision 5.b.1). Input to management option table.


Table 4.7.2. Faroe Plateau cod (subdivision 5.b.1). Management option table.

| MFDP VERSİN 1 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Run: Cod_farp |  |  |  |  |  |  |
| Index file 29/4-2016 |  |  |  |  |  |  |
| Time and date: 15:12 29/04/2016 |  |  |  |  |  |  |
| Fbar age range: 3-7 |  |  |  |  |  |  |
| 2016 |  |  |  |  |  |  |
| Biomass | SSB | FMult | FBar | Landings |  |  |
| 26625 | 22408 | 1.0000 | 0.4599 | 7514 |  |  |
| 201 |  |  |  |  | 201 |  |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 25745 | 20162 | 0.0000 | 0.0000 | 0 | 33196 | 26726 |
| . | 20162 | 0.1000 | 0.0460 | 869 | 32203 | 25774 |
| . | 20162 | 0.2000 | 0.0920 | 1700 | 31254 | 24864 |
| . | 20162 | 0.3000 | 0.1380 | 2494 | 30347 | 23995 |
| . | 20162 | 0.4000 | 0.1840 | 3254 | 29481 | 23166 |
| . | 20162 | 0.5000 | 0.2300 | 3980 | 28652 | 22374 |
| . | 20162 | 0.6000 | 0.2759 | 4676 | 27860 | 21617 |
| . | 20162 | 0.7000 | 0.3219 | 5341 | 27102 | 20894 |
| . | 20162 | 0.8000 | 0.3679 | 5978 | 26377 | 20203 |
| . | 20162 | 0.9000 | 0.4139 | 6587 | 25684 | 19542 |
| . | 20162 | 1.0000 | 0.4599 | 7170 | 25020 | 18911 |
| . | 20162 | 1.1000 | 0.5059 | 7729 | 24385 | 18307 |
| . | 20162 | 1.2000 | 0.5519 | 8265 | 23777 | 17730 |
| . | 20162 | 1.3000 | 0.5979 | 8777 | 23194 | 17178 |
| . | 20162 | 1.4000 | 0.6439 | 9269 | 22637 | 16650 |
| . | 20162 | 1.5000 | 0.6899 | 9740 | 22102 | 16144 |
| . | 20162 | 1.6000 | 0.7358 | 10192 | 21591 | 15661 |
| . | 20162 | 1.7000 | 0.7818 | 10625 | 21100 | 15198 |
| . | 20162 | 1.8000 | 0.8278 | 11040 | 20630 | 14755 |
| . | 20162 | 1.9000 | 0.8738 | 11439 | 20179 | 14330 |
| . | 20162 | 2.0000 | 0.9198 | 11821 | 19747 | 13924 |

Input units are thousands and kg -output in tonnes

Table 4.8.1. Faroe Plateau cod (subdivision 5.b.1). Input to yield-per-recruit calculations (long-term prediction).

|  | EXPL. | Weight | Prop |
| :---: | :---: | :---: | :---: |
|  | Pattern | AT AGE | MATURE |
|  |  |  |  |
|  | AVERAGE | AVERAGE | AVERAGE |
| AGE | $2002-2015$ | $1978-2015$ | $\mathbf{1 9 8 3 - 2 0 1 6}$ |
| NOT RESCALED |  |  |  |
| 2 | 0.103 | 1.044 | 0.06 |
| 3 | 0.328 | 1.571 | 0.58 |
| 4 | 0.452 | 2.296 | 0.84 |
| 5 | 0.572 | 3.097 | 0.94 |
| 6 | 0.681 | 3.903 | 0.98 |
| 7 | 0.735 | 4.999 | 0.99 |
| 8 | 0.725 | 6.254 | 1.00 |
| 9 | 0.988 | 7.820 | 1.00 |
| $10+$ | 0.988 | 9.676 | 1.00 |

Table 4.8.2. Faroe Plateau cod (subdivision 5.b.1). Output from yield-per-recruit calculations (longterm prediction).

| Reference point | F multiplier | Absolute F |
| :---: | :---: | :---: |
| $\mathrm{Fbar}^{\text {b }}$ (3-7) | 1.0000 | 0.5536 |
| $\mathrm{F}_{\text {max }}$ | 0.4576 | 0.2533 |
| $\mathrm{F}_{0.1}$ | 0.2112 | 0.1169 |
| $\mathrm{F}_{35 \%} \mathrm{SPR}$ | 0.3175 | 0.1758 |
| Flow | 0.1862 | 0.1031 |
| $\mathrm{F}_{\text {med }}$ | 0.6715 | 0.3717 |
| Fhigh | 1.6421 | 0.9091 |

### 4.19 Figures



Figure 4.2.1. Faroe Plateau cod (subdivision 5.b.1). Catch in numbers-at-age shown as catch curves.


Figure 4.2.2. Faroe Plateau cod (subdivision 5.b.1). Mean weight at age. The predicted weights are also shown.


Figure 4.2.3. Faroe Plateau cod (subdivision 5.b.1). Proportion mature at age as observed in spring groundfish survey. The predicted values are shown in grey.


Figure 4.2.4. Faroe Plateau cod (subdivision 5.b.1). Catch curves from spring groundfish survey.

Faroe Plateau cod


Figure 4.2.5. Faroe Plateau cod (subdivision 5.b.1). Stratified $\mathrm{kg} / \mathrm{hour}$ in spring and summer surveys (upper figure). The age 3+ biomass obtained from the assessment is also included as an index.

Figure 4.2.6. Faroe Plateau cod (subdivision 5.b.1). Catch curves from summer groundfish survey.


Figure 4.2.7. Faroe Plateau cod (subdivision 5.b.1). Standardized catch per unit of effort for pairtrawlers and longliners. The two surveys are shown as well.


Figure 4.2.8. Faroe Plateau cod (subdivision 5.b.1). Catch per unit of effort for small and large longliners compared with the fishable (age $3+$ ) biomass.


Figure 4.2.9. Faroe Plateau cod (subdivision 5.b.1). Catchability (cpue divided by age $3+$ biomass) for small and large longliners and pairtrawlers.

## Spring survey (shifted back to December)



Summer survey


Figure 4.6.1. Faroe Plateau cod (subdivision 5.b.1). Log catchability residuals for age 2-7 for spring (upper figure) and summer survey. The residuals for age 8 are not presented because some values were off scale. White bubbles indicate negative residuals.


Figure 4.6.2. Faroe Plateau cod (subdivision 5.b.1). Spawning-stock biomass (SSB) and recruitment (year class) vs. year (upper figure) and yield and fishing mortality vs. year. Points (white and grey) are taken from the short-term projections.


Figure 4.6.3. Faroe Plateau cod (subdivision 5.b.1). Different measures of fishing mortality: straight arithmetic average (Avg F), weighted by stock numbers (Nwtd), weighted by stock biomass (Bwtd) or weighted by catch (Cwtd).


Figure 4.6.4. Faroe Plateau cod (subdivision 5.b.1). Spawning stock - recruitment relationship. Years are shown at each data point.

Precautionary Approach Plot
Period 1959-2016


Figure 4.6.5. Faroe Plateau cod (subdivision 5.b.1). Spawning-stock biomass vs. fishing mortality.


Figure 4.6.6. Faroe Plateau cod (subdivision 5.b.1). Biomass (age 2+) obtained from the age-based assessment as well as from cpue of British trawlers and Faroe/Shetland vessels that were scaled to biomass. Prior to 1860 the export of dried cod was used as a basis to estimate the biomass also accounting for increased tendency to export dried cod during the period. The high estimates around 1850 are based on the assumption that the high values of dried cod export were due to an increased biomass alone. The lower estimates are based on the assumption that dried cod export increased, not only due to increased biomass, but also due to better possibilities to land fish during this period.


Figure 4.7.1. Faroe Plateau cod (subdivision 5.b.1). Predictions of the contribution of various year classes to the spawning-stock biomass in terminal year +1 (upper figure) and terminal year +2 (lower figure).


Figure 4.8.1. Faroe Plateau cod (subdivision 5.b.1). Yield-per-recruit and spawning-stock biomass (SSB) per recruit vs. fishing mortality (left figure). Landings and SSB versus Fbar (3-7) (right figure).


Figure 4.9.1. Faroe Plateau cod (subdivision 5.b.1). Results from the XSA retrospective analysis of fishing mortality (ages 3-7).


Figure 4.9.1. Faroe Plateau cod (subdivision 5.b.1). Results from the XSA retrospective analysis (continued). Recruitment-at-age 1 (upper figure) and at age 2.


Figure 4.9.1. Faroe Plateau cod (subdivision 5.b.1). Results from the XSA retrospective analysis (continued). Spawning-stock biomass (upper figure) and total-stock biomass.


Figure 4.9.2. Faroe Plateau cod (subdivision 5.b.1). Modelling cod recruitment in three steps. First, the catch-per-unit -effort of cod (C) for small boats operating close to land, as being indicative of the amount of cannibalistic cod. Second, the amount of cod (older than the recruiting cod) (B), as being indicative of e.g. the amount of schools to which recruiting cod can join and hide in. Third, the ratio between $B$ and $C$, as indicative of recruitment success. Fourth and fifth, a comparison with observed recruitment. Note that the model predicts that the recruitment in recent years is very poor.


Figure 4.10.1. Faroe Plateau cod (subdivision 5.b.1). Comparison between the results from the current assessment (Assm. 2016) and the assessment last year (Assm. 2015) for recruitment (upper left), fishing mortality (upper right), stock biomass (lower left) and spawning-stock biomass (lower right).

### 5.1 Stock description and management units

Haddock in Faroese Waters, i.e. ICES Subdivisions 5.b. 1 and 5.b. 2 and in the southern part of ICES Division 2.a, close to the border of Subdivision 5.b.1, are generally believed to belong to the same stock and are treated as one management unit named Faroe haddock. Haddock is distributed all over the Faroe Plateau and the Faroe Bank from shallow water down to more than 450 m . A more detailed description of haddock in Faroes waters is given in the stock annex. The spatial distribution of the haddock in the summer survey and in spring survey is shown in figure 5.9. The figure do clearly illustrate the drastic decrease in the stock biomass in recent years.

### 5.2 Scientific data

## Trends in landings and fisheries

Nominal landings of Faroe haddock increased very rapidly from only 4000 t in 1993 to 27000 t in 2003, but have declined drastically since and amounted in 2012 to only about 2600 t ; they have increased a bit to 3400 t in 2015. Most of the landings are taken from the Faroe Plateau; the 2015 landings from the Faroe Bank (Subdivision 5.b.2), where the area shallower than 200 m depths has been closed to almost all fishing since the fiscal year 2008-2009, amounted to only about 31 t (Tables 5.1 and 5.2). The cumulative landings by month are shown in Figure 5.2.

Faroese vessels have taken almost the entire catch since the late 1970s (Figure 5.1). Due to the dispute on mackerel quota share, there was no agreement on mutual fishery rights between the Faroe Islands and Norway and EU, respectively, since 2011 and therefore there was no fishery by those parties in 5.b in 2012 and 2013; in 2014 the parties happened to make an agreement again. The proportion of the Faroese landings taken by each fleet category since 1985 are shown in the annex. The longlines have taken most of the catches in recent years followed by the trawlers. This was also the case in 2015, where the share by longlines was $81 \%$ and that by trawlers $19 \%$ (Figure 5.3).

## Catch-at-age

Catch-at-age data were provided for fish taken by the Faroese fleets from 5.b. 1 and 5.b.2. The sampling intensity in 2015 is shown in Table 5.3 showing some decrease in intensity as compared to 2014. There is a need to increase the sampling level. Reasons for the inadequate sampling level are shortage of resources (people, money) but also that the total catches (and stock) are so small that it is difficult to obtain enough samples. From late 2011, a landing site has been established in Tórshavn close to the Marine Research Institute and it is the intention that technicians from the Institute will regularly be sampling the landings there; this will increase the sampling level in coming years. This has also turned out to be difficult of the above mentioned reasons but the outlook is very positive regarding raising enough money to hire a new technician to among other things do the sampling. The normal procedure has been to disaggregate samples from each fleet category by season (Jan-Apr, May-Aug and Sep-Dec) and then raise them by the corresponding catch proportions to give the annual catch-at-age in numbers for each fleet This year, all longliners were grouped into 2 fleets (larger and smaller than 100 GRT, respectively), and all trawlers were also grouped into 2 fleets
(larger and smaller than 1000 Hp , respectively). The longliner samples had to be treated by using 2 seasons only (Jan-Jun, Jul-Dec. The results are given in Table 5.3. No catch-at-age data were available from other nations (Norwegian longliners and British trawlers) and they were assumed to have the same age composition as the Faroese corresponding fleets. The most recent data were revised according to the final catch figures. The resulting total catch-at-age in numbers is given in Tables 5.4 and 5.5, and in Figure 5.4 the LN (catch-at-age in numbers) is shown since 1957.

In general the catch-at-age matrix in recent years appears consistent although from time to time some few very small year classes are disturbing this consistency, both in numbers and mean weights at age. The recent very small year classes need to be very carefully inspected when the $\mathrm{F}_{\mathrm{BAR}}$ is calculated. Also there are some problems with what ages should be included in the plus group; there are some periods where only a few fish are older than 9 years, and other periods with a quite substantial plus group $(10+)$. These problems have been addressed in former reports of this WG and will not be further dealt with here (See the 2005 NWWG report). Next year there will be a benchmark assessment of this stock, and all issues will be carefully investigated. No estimates of discards of haddock are available. However, since almost no quotas are used in the management of the fisheries on this stock, the incentive to discard in order to highgrade the catches should be low. The landings statistics is therefore regarded as being adequate for assessment purposes. The ban on discarding as stated in the law on fisheries should also - in theory - keep the discarding at a low level.

## Weight-at-age

Mean weight-at-age data are provided for the Faroese fishery (Table 5.5). Figure 5.5 shows the mean weights-at-age in the landings for age groups 2-7 since 1976. During this period, weights have shown cyclical changes. They were at a minimum in 20072009, but have increased again since then In the 3 latest years the weights have been fluctuated without a clear trend and a simple average of these years will be used in the short-term predictions (Figure 5.5). The mean weights at age in the stock are assumed equal to those in the landings.

## Maturity-at-age

Maturity-at-age data are available from the Faroese Spring Groundfish Surveys 19822016. The survey is carried out in February-March, so the maturity-at-age is determined just prior to the spawning of haddock in Faroese waters, mostly in April, and the determinations of the different maturity stages is relatively easy.
In order to reduce year-to-year effects due to possible inadequate sampling and at the same time allow for trends in the series, the routine by the WG has been to use a 3-year running average in the assessment. For the years prior to 1982, average maturity-atage from the surveys 1982-1995 was adopted (Table 5.6 and Figure 5.6).

### 5.3 Information from the fishing industry

There exists a considerable amount of data on fish size in the fishing industry. No such information was used directly in the 2016 assessment but catch per unit of effort for some selected fleets (logbook data) is used as an additional information on the status of the stock (see section 5.4.1.1).

### 5.4 Methods

This assessment is an update of the 2015 assessment, with exactly the same settings of the XSA. The only changes are minor revisions of recent landings according to revised data and corresponding revisions of the c@age input. All other input files (VPA) are the same except for the addition of the 2015data.

### 5.4.1 Tuning and estimates of fishing mortality

Commercial cpue series. Several commercial catch per unit of effort series are updated every year, but as discussed in previous reports of this WG they are not used directly for tuning of the VPA but as additional information on stock trends (for details see the stock annex). The age-aggregated cpue series for longliners and pairtrawlers are presented in Figure 5.7. In general the two series show the same trends although in some periods the two series are conflicting; this has been explained by variations in catchability of the longlines due to changes in productivity of the. Both series, however, indicate that the stock is very low. The longliner cpue's do not decrease as much as the trawler cpue's which in addition to the explanation given above may be attributed to the fact that in the management of the demersal Faroese stocks, large areas have been closed to trawling with the effect that when the haddock stock is small, the distribution of it is mainly outside the "trawl areas".

In order to illustrate stock biomass further back in time, historical cpue series from British trawlers have been used together with the 2015 assessment. The method is described in WD12, "Faroe haddock biomass 1914-56". The results are given in Table 5.17, and in section 02 of this report. The biomass of Faroe haddock was estimated back to 1914 by scaling cpue values to the biomass obtained from the stock assessment. There was an overlap between cpue values from Aberdeen trawlers 1914-1959 and the age based assessment 1957-2015 by three years (1957-1959). Cpue values for English steam trawlers from 1922 to 1976 (with gaps) confirmed that the former overlap of three years was sufficient to provide a scaling of Aberdeen trawler cpue back to 1914 (Table 5.17). The table shows that the low biomass since 2006 has been unprecedented the last century.
Fisheries independent cpue series. Two annual groundfish surveys are available, one carried out in February-March since 1982 ( 100 stations per year down to 500 m depth), and the other in August-September since 1996 (200 stations per year down to 500 m depth). Biomass estimates ( $\mathrm{kg} / \mathrm{hour}$ ) are available for both series since they were initiated (Figure 5.8). The main trends from the surveys are the same but the summer survey indicates a considerably more depleted stock in recent years than the summer survey; both surveys indicate a slow increase in recent years. Age disaggregated data are available for the whole summer series, but due to problems with the database (see earlier reports), age disaggregated data for the spring survey are only available since 1994. The calculation of indices at age is based on age-length keys with a smoother applied. This is a useful method but, some artefacts may be introduced because the smoothing can assign wrong ages to some lengths, especially for the youngest and oldest specimen. As in recent years, the length distributions have been used more directly for calculation of indices at age (ages 0-2), since these ages have length distributions almost without overlap. LN (numbers-at-age) for the surveys are presented in Figures 5.10-5.11. Further analyses of the performances of the two series are shown in the stock annex. In general there is a good relationship between the indices for one year class in two successive years.

A SPALY (same procedure as last year) run, with the same settings of the XSA as in 2015 (tuned with the two surveys combined) (Table 5.8), with 2016 data included and some minor revisions of recent catch figures, gave in general similar results as last year (Table 5.9), although this year's assessment indicates that the 2015 assessment underestimated the 2014 recruitment by $42 \%$ ( 2.6 million vs. 4.4 .5 million, which still is among the lowest on record), underestimated the fishing mortality in 2014 by $10 \%$ ( 0.26 vs. 0.29 ) and overestimated the 2014 total and spawning-stock biomasses by $7 \%$ and $27 \%$, respectively ( 20 and 18 thous. t vs. 18 and 14 thous. t

The $\log \mathrm{q}$ residuals for the two surveys are shown in Figure 5.12
The retrospective analysis of fishing mortality, recruitment and spawning-stock biomass of this XSA is shown in Figure 5.13. The retrospective pattern of the fishing mortality is hampered by strange values of some small poorly sampled year classes which in some years are included in the Fbar reference ages and consequently they will create problems for estimation of the stock (see the 2005 NWWG report); this is not a problem for the time being but the development of recent small year classes are being carefully inspected.

It has been questioned if a rather heavy shrinkage of 0.5 is the most appropriate to a stock like Faroe haddock where biological parameters and fishing mortality (catchability) are closely linked to productivity changes in the ecosystem. In order to investigate the possible effect of the shrinkage, the 2010 NWWG carried out an exploratory XSA without shrinkage (Shr. 2.0). Based on that it was concluded to continue with a shrinkage of 0.5 and this shrinkage was also applied this year.

Results. The fishing mortalities from the final XSA run are given in Table 5.9 and in Figure 5.14. The fishing mortality was high (around 0.6) in the 1950s and early 1960s but declined to around 0.2 from 1965-1975. Since then, fishing mortality has usually been low, the exceptions are peaks in 1977, 1982, 1997-1999 and 2003-2006. They occur near the end of relatively high catch periods and some of the highest values ( $0.32-0.45$ ) are nearly certainly an artefact of the unweighted fishing mortality. Exploitation ratio (Yield/Biomass) is a bit more stable and may be used to indicate the level of fishing mortality.

### 5.5 Reference points

The yield- and spawning-stock biomass per recruit (age 2) based on the long-term data are shown in Table 5.16 and Figure 5.16. From Table 5.16, F med, $_{\text {, and }}$ Fhigh were calculated at 0.22 and 0.82 , respectively. The $F_{\max }$ of 0.6 should not be used since it is very poorly determined due to the flat YPR curve. Fo.1 is estimated at 0.2. The F35\%SPR was estimated at 0.23.

The precautionary reference fishing mortalities were set in 1998 by ACFM with $\mathrm{F}_{\mathrm{pa}}$ as the $\mathbf{F}_{\text {med }}$ value of 0.25 and Flim two standard deviations above $\mathbf{F}_{\text {pa }}$ equal to 0.40 . The precautionary reference spawning-stock biomass levels were changed by ACFM in 2007. $B_{\text {lim }}$ was set at $22000 t\left(B_{\text {loss }}\right)$ and $B_{p a}$ at $35000 t$ based on the formula $B_{p a}=B_{\lim } e^{1.645 \sigma}$, assuming a $\sigma$ of about 0.3 to account for the uncertainties in the assessment.

The working group in 2012 investigated possible candidates for Fmš. Based on Medium -term projections, Medium-term projections the NWWG suggested, that FMSY preliminary could be set at 0.25 and the MSY Btrigger at 35 thous. $t$ (same as $B_{p a}$ ) These values were accepted by ACOM. Some further analyses have indicated that these
values are acceptable, but it is anticipated that further work will be untertaken in connection with the next benchmark assessment. See the stock annex for more details.

### 5.6 State of the stock - historical and compared to what is now.

The stock size in numbers is given in Table 5.11 and a summary of the VPA with the biomass estimates is given in Table 5.12 and in Figure 5.14. According to this assessment, the period up to the mid 1970s was characterized by relative high and stable landings, recruitment and spawning-stock biomass and the stock was able to withstand relatively high fishing mortalities. Since then the spawning-stock biomass has shown large fluctuations due to cyclical changes in recruitment, growth and maturity (Figures 5.5 and 5.6). The fishing mortality does not seem to be the decisive factor in this development since it most of the period has fluctuated around the Fmsy and $\mathrm{F}_{\mathrm{pa}}$. It must though be remembered that the characteristics of the stock in recent decades with long periods of poor recruitment make it less resilient to high fishing mortality.

The most recent increase in the spawning stock is due to new strong year classes entering the stock of which the 1999 year class is the highest on record (103 million at age 2). Also the YC's from 2000 and 2001 are estimated well above average and the 2002 YC above average, but the more recent YC's are all estimated to be very small except the 2009 YC, which is estimated to be slightly above the half of the average for the whole series back to 1957 and the 2010 and 2014 YC's, which are estimated somewhat higher than the other small year classes. Fishing mortality has been relatively high since 2003, highest when the stock was large leading to large variability of catches. Currently fishing mortality is estimated close to $\mathrm{F}_{\text {msy }}$ (0.25).

### 5.7 Short-term forecast

## Input data

The input data for the short-term predictions are estimated in accordance with the procedures last year and explained in Tables 5.12-13. The YC 2016 at age 2 in 2018 is estimated as the geometric mean of the 2 -year-olds since 2005 . This procedure was introduced in 2011. All available information suggests that using the recent short series with poor recruitment is more appropriate than the longer period used in the past. However, the choice of recruitment in 2018 has little effect on the short-term prediction.

## Results

Although the allocated number of fishing days for the fishing year 2015-2016was reduced for some fleets as compared to the year before (see section 2), it should not be unrealistic to assume fishing mortalities in 2016 as the average of some recent years, here the average of F (2013-2015), since not all allocated days were actually used; however, possible changes in the catchability of the fleets (which seems to be linked to productivity changes in the environment) could undermine this assumption; price differences between cod and haddock may also influence this assumption. The landings in 2016 are then predicted to be about 4000 t , and continuing with this fishing mortality will result in 2017 landings of about 5300 t (Table 5.15). The SSB will increase to 20000 t in 2016, and increase further in 2017 and 2018 to 24000 t and 41000 t , respectively. This prediction should however be treated with great care since most of the increase is based on number of 1 year old in the 2016 spring survey. The results of the short-term prediction are shown in Table 5.16 and in Figure 5.14. The contribution (\%) by year classes to the age composition of the predicted 2016 and 2017 SSB's is shown in Figure
5.17. It should be noted that young $\mathrm{YC}^{\prime}$ 's which not have really entered the fishery in 2015/16, will contribute by a large proportion of the SSB in 2017/18.

### 5.8 Medium term forecasts and yield-per-recruit

No medium term projections were made this year; however, the 2013 projections, which were the basis for suggested MSY reference points, are presented in the stock annex.

The input data for the long-term yield and spawning-stock biomass (yield-per-recruit calculations) are listed in Table 5.15. Mean weights-at-age (stock and catch) are averages for the 1977-2015 period. The maturity o-gives are averages for the years 19822015. The exploitation pattern is the same as in the short-term prediction.

The results are given in Table 5.16, in Figure 5.16 and under Reference points (section 5.5).

### 5.9 Uncertainties in assessment and forecast

Retrospective analyses indicate periods with tendencies to overestimate spawningstock biomass and underestimate fishing mortality and vice versa. Similar things can be seen with the recruitment. This year's assessment indicates that the 2015 assessment underestimated the 2014 recruitment by $42 \%$ ( 2.6 million vs. 4.5 million, which still is among the lowest on record), underestimated the fishing mortality in 2014 by 10\% ( 0.29 vs. 0.26 ) and overestimated the 2014 total- and spawning-stock biomasses by $7 \%$ and $27 \%$, respectively ( 20 and 18 thous. t vs. 19 and 14 thous. t ), see text table below..

Recruitment estimates from surveys are not very consistent for small cohorts...
The sampling of the catches for length measurements, otolith readings and lengthweight relationships has decreased somewhat compared to 2015. Although it is regarded to be adequate for the assessment, there is a need to improve it again (see 5.2).

### 5.10 Comparison with previous assessment and forecast

As explained previously in the report, this assessment is an update of the 2015 assessment. The only changes are minor revisions of recent landings according to revised data and corresponding revisions of the c@age input. All other input files (VPA and tuning fleets) are the same except for the addition of the 2015 data.
Following differences in the 2014 estimates were observed as compared to last year (see text above):

## Comparisons between 2015 and 2016 assessment of 2014 data

 The year of comparison is 2014|  | R at age 2 <br> (thousands) | Total B <br> (tonnes) | SSB <br> (tonnes) | Landings <br> (tonnes) | F (3-7) |
| :--- | :---: | :---: | :---: | :---: | :---: |
| 2015 spaly | 2596 | 19643 | 17931 | 2950 | 0.2595 |
| 2016 spaly | 4513 | 18411 | 14083 | 3276 | 0.2876 |
| \%-change | 42 | -7 | -27 | 10 | 10 |

### 5.11 Management plans and evaluations

There is no explicit management plan for this stock. A management system based on number of fishing days, closed areas and other technical measures was introduced in 1996 with the purpose to ensuring sustainable fisheries. There has been some work with establishing a management plan with a harvest control role for cod, haddock and saithe including a recovery plan, but the proposal has not yet been officially accepted. There is ongoing work with a revision on most aspect of the fisheries legislation. See overview in section 2 for details.

### 5.12 Management considerations

Management of fisheries on haddock also needs to take into account measures for cod and saithe.

### 5.13 Ecosystem considerations

Since on average about $80 \%$ of the catches are taken by longlines and the remaining by trawls, effects of the haddock fishery on the bottom is moderate.

### 5.14 Regulations and their effects

As explained in the overview (section 2), the fishery for haddock in $5 . b$ is regulated through a maximum number of allocated fishing days, gear specifications, closed areas during spawning times, closed areas for longlining close to land and large areas closed to trawling. As a consequence, around $80 \%$ of the haddock landings derive from longline fisheries. Since the minimum mesh size in the trawls (codend) is 145 mm , the trawl catches consist of fewer small fish than the longline fisheries. Other nations fishing in Faroese waters are regulated by TACs obtained during bilateral negotiations; their total landings are minimal, however, and in 2011-2013 no agreement could be made between the Faroe Islands and EU and Norway, respectively, due to the dispute on mackerel quota sharing. In 2014 and 2015, however, the parties managed to get an agreement in place again. Discarding of haddock is considered minimal and there is a ban to discarding.

### 5.15 Changes in fishing technology and fishing patterns

See section 2.

### 5.16 Changes in the environment

See section 2.

### 5.17 Tables

Table 5.1 Faroe Plateau (Sub-division 5.b.1) HADDOCK. Nominal catches (tonnes) by countries 2000-2015 and Working Group estimates in 5.b.

| Country | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | $2015{ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | $13,620^{8}$ | $13,457{ }^{8}$ | 20,776 ${ }^{6}$ | 21,615 | 18,995 | 18,172 | 15,600 | 11,689 | 6,728 | 4,895 | 4,932 | 3,350 | 2,490 | 2,877 | 2,756 | 2,910 |
| France1 | 6 | $8^{7}$ | 2 | 4 | $1^{5}$ | + | $12^{5}$ | $4^{5}$ | $3^{5}$ | $2^{5}$ | $1{ }^{7}$ | 3 |  |  |  | + |
| Germany | 1 | 2 | 6 | 1 | 6 |  | 1 |  |  |  |  |  |  |  |  |  |
| Greenland | $22^{6}$ | $0^{6}$ | $4^{4}$ |  |  |  | 1 | $9{ }^{4}$ |  | $6{ }^{4}$ | $12^{6}$ | + | $1{ }^{4}$ |  |  |  |
| Iceland |  |  | 4 |  |  |  |  |  |  |  |  |  | 2 | $26^{4}$ |  |  |
| Norway | 355 | $257{ }^{2}$ | 227 | 265 | 229 | 212 | 57 | 61 | 26 | 8 | 5 |  |  |  | 2 | 5 |
| Russia |  |  |  |  | 16 |  |  |  | 10 |  |  |  |  |  |  |  |
| Spain |  |  |  |  | 49 |  |  |  |  |  |  |  |  |  |  |  |
| UK (Engl. and Wales) | $19^{7}$ | $4^{7}$ | $11^{5}$ | 14 | 8 | 1 | 1 |  |  |  |  |  |  |  |  |  |
| UK (Scotland) 5 |  |  |  | 185 | 186 | 126 | 106 | 35 | 60 | 64 |  |  |  |  |  |  |
| United Kingdom |  |  |  |  |  |  |  |  |  |  | $73{ }^{4}$ |  |  |  | 350 | 449 |
| Total | 14,023 | 13,728 | 21,030 | 22,084 | 19,490 | 18,511 | 15,778 | 11,798 | 6,827 | 4,975 | 5,023 | 3,353 | 2,493 | 2,903 | 3,130 | 3,364 |
| Used in the assessment in 5.b. | 15,821 | 15,890 | 24,933 | 27,072 | 23,101 | 20,455 | 17,154 | 12,631 | 7,388 | 5,197 | 5,202 | 3,540 | 2,634 | 2,950 | 3,276 | 3,395 |

1) Including catches from Sub-division 5.b.2. Quantity unknown 1989-1991, 1993 and 1995-2001
2) Preliminary data
3) From 1983 to 1996 catches included in Sub-division 5.b.2.
4) Reported as Division 5. b, to the Faroese coastal guard service.
5) Reported as
6) Reported as Division 5.b.
7) Includes Faroese landings reported to the NWWG by the Faroe Marine Research Institute

Table 5.2 Faroe Bank (Sub-division 5.b.2) HADDOCK. Nominal catches (tonnes) by countries, 2000-2015.

| Country | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | $2015{ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Island | 1,565 ${ }^{\text {s }}$ | 1,948 | 3,698 | 4,934 | 3,594 | 2,444 | 1,375 | 810 | 556 | 192 | 178 | 194 | 141 | 47 | 71 | 30 |
| France1 |  |  |  |  |  | + |  |  |  |  |  |  |  |  |  |  |
| Norway | 48 | 66 | 28 | 54 | 17 | 45 | 1 | 8 |  | 3 | 1 |  |  |  | 1 | 1 |
| UK (Engl. and | les) | : | : | : | : | 1 | 4 |  |  |  |  |  |  |  |  |  |
| UK (Scotlar | 185 | 148 | 177 | 4 | : | 1 | 4 | 15 | 5 | 27 | 33 |  |  |  | 74 |  |
| Total | 1,798 | 2,162 | 3,903 | 4,988 | 3,611 | 1,944 | 1,376 | 833 | 561 | 222 | 212 | 194 | 141 | $47^{\prime \prime}$ | $146^{\prime \prime}$ | 31 |

1) Catches included in Sub-division 5.b.1
2) Provisional data
3) From 1983 to 1996 includes also catches taken in Sub-division 5.b.1. (see Table 2.4.1)
4) Reported as Division 5.b.
5) Provided by the NWWG

Table 5.3
Catch at age 2015

| Age | 5.b LLiners $<100 \mathrm{GRT}$ | $\begin{gathered} \hline \text { 5.b } \\ \text { LLiners } \\ >100 \mathrm{GRT} \\ \hline \end{gathered}$ | 5.b Trawl $<1000 \mathrm{HP}$ |  <br>  <br>  <br> Trawl <br> $>$ <br> $>$ | 5.b Regulator | $5 . b$ <br> All Faroese <br> fleets | 5.b Foreign Trawlers | 5.b Foreign Lliners | 5.b Total All fleets |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 1 |
| 2 | 260 | 81 | 4 | 16 | 2 | 362 | 24 | 0 | 384 |
| 3 | 633 | 211 | 71 | 92 | 0 | 1007 | 137 | 1 | 1144 |
| 4 | 87 | 59 | 43 | 52 | 2 | 242 | 77 | 0 | 318 |
| 5 | 212 | 193 | 42 | 47 | 3 | 493 | 70 | 1 | 560 |
| 6 | 114 | 116 | 23 | 28 | 2 | 281 | 42 | 1 | 322 |
| 7 | 12 | 17 | 4 | 6 | 0 | 40 | 10 | 0 | 50 |
| 8 | 4 | 11 | 2 | 4 | 0 | 21 | 5 | 0 | 27 |
| 9 | 3 | 14 | 1 | 2 | 0 | 20 | 3 | 0 | 23 |
| 10 | 4 | 3 | 1 | 1 | 0 | 10 | 2 | 0 | 11 |
| 11 | 1 | 3 | 0 | 1 | 0 | 4 | 1 | 0 | 5 |
| 12 | 0 | 0 | 0 | 1 | 0 | 1 | 1 | 0 | 2 |
| 13 | 1 | 1 | 0 | 0 | 0 | 2 | 0 | 0 | 2 |
| 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total no. | 1333 | 709 | 192 | 251 | 16 | 2485 | 373 | 4 | 2847 |
| Catch, t. | 1296 | 896 | 201 | 273 | 17 | 2665 | 405 | 5 | 3058 |

Notes: $\quad$ Numbers in $1000^{\circ}$
Catch, gutted weight in tonnes
Others includes netters, jiggers, other small categories and catches not otherwise accounted for
LLiners = Longliners OB.trawl. = Otterboard tra^ Pair Trawl. = Pair trawlers

| $\begin{gathered} \hline \text { Comm. } \\ \text { Sampling } \\ 2015 \\ \hline \end{gathered}$ | 5.b LLiners $<100 \mathrm{GRT}$ | 5.b LLiners $>100 \mathrm{GRT}$ | 5.b Trawl $<1000 \mathrm{HP}$ | 5.b Trawl $<1000 \mathrm{HP}$ | 5.b Regulator | $5 . b$ <br> All Faroese <br> Fleets | 5.b Foreign Trawlers | 5.b <br> Foreign Lliners | $\begin{gathered} \text { 5.b } \\ \text { Total } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. samples | 7 | 14 | 9 | 34 | 0 | 64 | 0 | 0 | 73 |
| No. lengths | 1525 | 2947 | 1599 | 7476 | 0 | 13547 | 0 | 0 | 16942 |
| No. weights | 1525 | 2947 | 1599 | 7476 | 0 | 13547 | 0 | 0 | 16942 |
| No. ages | 140 | 300 | 159 | 589 | 0 | 1188 | 0 | 0 | 1379 |

As compared to 2014, the sampling in 2015 was:
no samples $-7 \%$, no of lengths $-8 \%$, no of weights $5 \%$, no of otoliths $-4 \%$.

Table 5.4 Faroe haddock. Catch number-at-age


|  | Table 1 | Catch numbers-at-age |  |  |  |  | Numbers*10**-3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984, | YEAR, 1985, | 1976, | 1977, | 1978, | 1979, | 1980, | 1981, | 1982, | 1983, |
| AGE |  |  |  |  |  |  |  |  |  |
|  | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , |
| 0, | 0 , |  |  |  |  |  |  |  |  |
|  | 1, | 40, | 0 , | 0, | 1, | 0 , | 0 , | 0 , | 0 , |
| 25, | 0 , |  |  |  |  |  |  |  |  |
|  | 2, | 4396, | 255, | 32, | 1, | 143, | 74, | 539, | 441, |
| 1195, | 985, |  |  |  |  |  |  |  |  |
|  | 3 , | 7858, | 4039, | 1022, | 1162, | 58, | 455, | 934, | 1969, |
| 1561, | 4553, |  |  |  |  |  |  |  |  |
|  | 4, 2196 | 6798, | 5168, | 4248, | 1755, | 3724, | 202, | 784, | 383, |
| 2462 , | 5, | 1251, | 4918, | 4054, | 3343, | 2583, | 2586, | 298, | 422, |
| 147, | 1242, |  |  |  |  |  |  |  |  |
|  | 6, | 1189, | 2128, | 1841, | 1851, | 2496, | 1354, | 2182, | 93, |
| 234, | 169, |  |  |  |  |  |  |  |  |
|  | 7, | 298, | 946, | 717, | 772, | 1568, | 1559, | 973, | 1444, |
| 42, | 91, |  |  |  |  |  |  |  |  |
|  | 8, 61, | 720, | 443, | 635, | 212, | 660, | 608, | 1166, | 740, |
| 861, | 9, | 258, | 731, | 243, | 155, | 99, | 177, | 1283, | 947, |
| 388, | 503, |  |  |  |  |  |  |  |  |
|  | +gp, | 318, | 855, | 312, | 74, | 86, | 36, | 214, | 795, |
| 968, | 973, |  |  |  |  |  |  |  |  |
|  | TOTALNUM, | 23126, | 19483, | 13104, | 9326, | 11417, | 7051, | 8373, | 7234, |
| 7883, | 10773, |  |  |  |  |  |  |  |  |
|  | TONSLAND, | 26211, | 25555, | 19200, | 12424, | 15016, | 12233, | 11937, | 12894, |
| 12378, | , 15143, |  |  |  |  |  |  |  |  |
|  | SOPCOF \%, | 107, | 98, | 99, | 104, | 100, | 109, | 92, | 106, |
| 106, | 106, |  |  |  |  |  |  |  |  |

Table 5.4 Faroe haddock. Catch number-at-age (cont.)



## Table 5.5 Faroe haddock. Catch weight-at-age.




Table 5.5 Faroe haddock. Catch weight-at-age (cont.).


| $1.8430, \quad 7$ | 2.3510, | 2.3400, | 2.5560, | 2.4560, | 1.8930, | 2.1190, | 2.3010, | 2.0910, | 1.8700, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2.0610, $2,2.4690,2.4750,2.5720,2.6580,2.8210,2.3730,2.3700,2.3010,2.4380$, |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| 9, | 2.7770, | 2.5010, | 2.4520, | 2.5980, | 3.7490, | 2.7500, | 2.6260, | 2.4060, | 2.3570, |
| 2.2630, |  |  |  |  |  |  |  |  |  |
| +gp, | 2.5820, | 2.6760, | 2.7530, | 2.9530, | 3.1960, | 3.9660 , | 3.1300, | 2.5350, | 2.4170, |
| 2.5790, |  |  |  |  |  |  |  |  |  |
| SOPCOFAC, $.9988,$ | 1.0043, | 1.0250, | 1.0106, | . 9973 , | 1.0349, | . 9960 , | 1.0010, | 1.0049, | .9929, |
| Table 2 | Catch | weights | age (kg |  |  |  |  |  |  |
| 2015, YEAR, | 2006, | 2007, | 2008, | 2009, | 2010, | 2011, | 2012, | 2013, | 2014, |
| AGE |  |  |  |  |  |  |  |  |  |
| . $0000, .0000, .0000, .0000, .0000, .0000, .0000, .0000, .0000, .0000$, |  |  |  |  |  |  |  |  |  |
| .4240, 1, .0000, .0000, .4910, .0000, .0000, .0000, .0000, .0000, .0000, |  |  |  |  |  |  |  |  |  |
| .5330, 2, .4750, .6280, .6360, .4820, .6920, .5530, .6190, .5760, .5470, |  |  |  |  |  |  |  |  |  |
| .8890, 3, .6010, .6690, .7540, .7340, .8700, .8150, .7860, .8300, .9020, |  |  |  |  |  |  |  |  |  |
| 1.3530, 4, .7680, .8590, .8600, .9850, 1.1490, 1.0860, 1.0690, 1.1490, 1.1650, |  |  |  |  |  |  |  |  |  |
| 1.6400, 5, .9110, .9690, .9910, 1.1300, 1.3080, 1.3030, 1.4050, 1.4650, 1.3540, |  |  |  |  |  |  |  |  |  |
| 1.7290, 6, 1.1260, 1.0600, 1.0820, 1.2640, 1.3860, 1.3870, 1.6160, 1.7100, 1.6930, |  |  |  |  |  |  |  |  |  |
| 2.4240, 7, 1.3740, 1.2450, 1.1510, 1.3570, 1.4290, 1.4690, 1.6560, 1.8270, 1.8410, |  |  |  |  |  |  |  |  |  |
| 2.0030, 8, 2.1580, 1.4750, 1.3790, 1.5450, 1.5680, 1.5380, 1.6750, 1.8860, 1.8720, |  |  |  |  |  |  |  |  |  |
| 2.2180, 9, 2.2110, 2.2660, 1.7270, 1.7920, 1.7400, 1.7020, 1.7270, 1.8560, 1.8560, |  |  |  |  |  |  |  |  |  |
| $2.3020,{ }^{+g p,}$ | 2.5690, | 2.2560, | 2.4350, | 2.1540, | 1.8410, | 1.8620, | 1.9050, | 2.0850, | 1.8230, |
| SOPCOFAC, $.9994,$ | . 9987 , | .9999, | 1.0065, | .9955, | 1.0076, | 1.0060, | 1.0190, | 1.0077, | 1.0112, |

Table 5.6Faroe haddock. Proportion mature-at-age.

1.0000,

YEAR, 1966, 1967, 1968, 1969, 1970, 1971, 1972, 1973,

| 1974, | $\begin{aligned} & \text { YEAR, } \\ & 1975, \end{aligned}$ | 1966, | 1967, | 1968, | 1969, | 1970, | 1971, | 1972, | 1973, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |
|  | 0 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , |
| . 0000 , | . 0000 , |  |  |  |  |  |  |  |  |
|  | 1, | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , |
| . 0000 , | $\begin{aligned} & .0000, \\ & 2, \end{aligned}$ | . 0600 , | . 0600 , | . 0600 , | . 0600 , | . 0600 , | . 0600 , | . 0600, | . 0600 , |
| . 0600 , | $.0600$ |  |  |  |  |  |  |  |  |
|  | $3,$ | . 4800 , | . 4800 , | . 4800 , | . 4800 , | . 4800, | . 4800 , | . 4800, | . 4800 , |
| . 4800, | $\begin{aligned} & .4800, \\ & 4, \end{aligned}$ | . 9100 , | . 9100 , | . 9100 , | . 9100, | .9100, | . 9100 , | . 9100 , | . 9100, |
| . 9100, | $\begin{aligned} & .9100, \\ & 5, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | $\begin{aligned} & 1.0000, \\ & 6, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | $\begin{aligned} & 1.0000, \\ & 7, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | $\begin{aligned} & 1.0000, \\ & 8, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | $\begin{aligned} & 1.0000, \\ & 9, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | 1.0000, |  |  |  |  |  |  |  |  |
| $1.0000,$ | $\begin{aligned} & +\mathrm{gp}, \\ & \quad 1.0000, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |



| .0800, |  | . 0600 , | .0600, | . 0600 , | . 0600 , | . 0600 , | .0600, | .0800, | .0800, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3, | . 4800, | . 4800 , | . 4800 , | . 4800 , | . 4800 , | . 4800 , | . 6200, | .6200, |
| . 7600 , | $4$ | . 9100, | . 9100, | . 9100, | . 9100, | . 9100, | . 9100, | .8900, | .8900, |
| . 9800 , | $.9600,$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | $\begin{aligned} & 1.0000, \\ & 6, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | $\begin{aligned} & 1.0000, \\ & 7, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, 1.0000, | $\begin{aligned} & 1.0000, \\ & 8, \\ & 1.0000, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | $\begin{aligned} & 9, \\ & 1.0000, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | $\begin{aligned} & \mathrm{gp}, \\ & 1.0000, \end{aligned}$ | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |

Table 5.6 Faroe haddock. Proportion mature-at-age (cont.).

| 1995, | Table 5 Proportion mature at age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR, | 1986, | 1987, | 1988, | 1989, | 1990, | 1991, | 1992, | 1993, | 1994, |
|  | AGE |  |  |  |  |  |  |  |  |  |
| . 0000 , | 0 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , |
| $.0000$ | 1, | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , |
| . 0300, | 2, | . 0300, | . 0500 , | . 0500, | . 0200, | . 0800 , | . 1600, | . 1800, | .1100, | . 0500, |
| . 4700 , | 3 , | . 4300, | . 3200, | . 2400 , | . 2200, | . 3700 , | . 5800 , | .6500, | . 5000, | . 4200, |
| . 9100 , | 4, | . 9500 , | . 9100, | . 8900, | . 8700 , | . 9000 , | . 9300 , | . 9100, | . 8500, | . 8600 , |
| . 9600 , | 5, | . 9900 , | . 9800 , | . 9800 , | . 9900 , | 1.0000, | 1.0000, | 1.0000, | . 9700 , | . 9600 , |
| . 9900 , | 6, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | . 9900 , | . 9900 , |
| 1.0000, | 7, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | 8, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | 9, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 1.0000, | +gp, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |




Table 5.7. Faroe haddock. 2016 tuning file.


| 100 | 5798.80 | 6022.70 | 7742.00 | 6165.00 | 4565.90 | 4912.80 | 238.60 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 100 | 705.50 | 6284.80 | 1574.60 | 4457.00 | 3250.40 | 3267.40 | 1577.20 |
| 100 | 1191.70 | 1873.30 | 4202.40 | 1008.90 | 3511.30 | 3712.50 | 2875.00 |
| 100 | 667.90 | 2182.60 | 820.20 | 1694.90 | 599.50 | 1665.00 | 1463.80 |
| 100 | 4119.00 | 2079.00 | 1125.10 | 405.90 | 916.80 | 371.50 | 924.90 |
| 100 | 6945.00 | 4655.30 | 638.10 | 418.70 | 196.20 | 280.20 | 265.90 |
| 100 | 101.10 | 6320.00 | 1865.90 | 449.30 | 260.30 | 212.60 | 244.60 |
| 100 | 420.00 | 367.60 | 4957.20 | 908.00 | 227.80 | 142.50 | 293.30 |
| 100 | 3419.90 | 1232.21 | 302.60 | 4022.40 | 619.60 | 120.30 | 103.78 |
| 100 | 3542.60 | 4099.30 | 869.80 | 930.30 | 2238.40 | 270.20 | 90.30 |
| 100 | 1545.00 | 3327.70 | 4123.00 | 1086.10 | 2026.30 | 1296.40 | 184.10 |
| 100 | 12458.90 | 4441.90 | 2487.80 | 1332.90 | 263.00 | 428.50 | 107.00 |

Table 5.8Faroe haddock 2016 xsa.


## Table 5.8 Faroe haddock 2016 xsa (cont.)

XSA population numbers (Thousands)

| $\begin{aligned} & \text { YEAR , } \\ & 8, \end{aligned}$ | 9, 0, | 1, | AGE $2,$ | 3 , | 4, | 5, | 6 , | 7, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 3.97E+03, | 3.77E+03, | 7.59E+03, | $6.83 E+03$ | 1.68E+04, | 1.75E+04, | 1.43E+04, | 7.20E+03, |
| 4.91E+02, | 1.51E+02, |  |  |  |  |  |  |  |
| 2007 | 3.43E+03, | 3.25E+03, | 3.09E+03, | $5.99 \mathrm{E}+03$ | 5.19E+03, | 1.15E+04, | 1.08E+04, | $6.83 \mathrm{E}+03$, |
| 2.93E+03, | 2.79E+02, |  |  |  |  |  |  |  |
| 2008 | $6.30 \mathrm{E}+03$, | 2.81E+03, | 2.66E+03, | $2.46 \mathrm{E}+03$ | 4.01E+03, | $3.75 \mathrm{E}+03$, | $6.91 \mathrm{E}+03$, | 5.82E+03, |
| $3.10 \mathrm{E}+03$, | 1.39E+03, |  |  |  |  |  |  |  |
| 2009 | 1.74E+04, | $5.16 \mathrm{E}+03$ | 2.29E+03, | 2.12E+03, | 1.83E+03, | $2.45 \mathrm{E}+03$ | 2.69E+03, | 4.33E+03, |
| 3.22E+03, | 1.40E+03, |  |  |  |  |  |  |  |
| 2010 | $6.73 \mathrm{E}+03$, | 1.42E+04, | 4.22E+03, | 1.85E+03, | 1.44E+03, | 1.13E+03, | 1.51E+03, | 1.74E+03, |
| 2.52E+03, | 1.97E+03, |  |  |  |  |  |  |  |
| 2011 | $4.10 \mathrm{E}+03$, | 5.51E+03, | 1.17E+04, | 3.11E+03, | 1.12E+03, | $7.90 \mathrm{E}+02$ | $6.75 E+02$, | $7.97 \mathrm{E}+02$, |
| 9.22E+02, | 1.41E+03, |  |  |  |  |  |  |  |
| 2012 | 1.31E+04, | $3.35 E+03$ | 4.51E+03, | 9.39E+03, | 1.84E+03, | $6.20 \mathrm{E}+02$ | 4.68E+02, | $3.84 \mathrm{E}+02$, |
| $3.99 \mathrm{E}+02$, | 4.36E+02, |  |  |  |  |  |  |  |
| 2013 | 1.51E+04, | 1.07E+04, | 2.75E+03, | 3.69E+03, | $6.82 \mathrm{E}+03$, | 1.05E+03, | $3.67 \mathrm{E}+02$, | $2.80 \mathrm{E}+02$, |
| 2.03E+02, | 2.41E+02, |  |  |  |  |  |  |  |
| 2014 | 1.02E+04, | 1.24E+04, | 8.80E+03, | 2.17E+03, | 2.56E+03, | $4.57 \mathrm{E}+03$, | $6.58 \mathrm{E}+02$, | $2.14 \mathrm{E}+02$, |
| 1.58E+02, | 8.67E+01, |  |  |  |  |  |  |  |
| 2015 | $6.24 \mathrm{E}+04$, | 8.34E+03, | 1.01E+04, | $6.99 \mathrm{E}+03$ | 1.42E+03, | 1.51E+03, | 2.71E+03, | 4.46E+02, |
| 1.20E+02, | 1.01E+02, |  |  |  |  |  |  |  |

Estimated population abundance at 1st Jan 2016
$0.00 \mathrm{E}+00,5.11 \mathrm{E}+04,6.83 \mathrm{E}+03,7.94 \mathrm{E}+03,4.69 \mathrm{E}+03,8.77 \mathrm{E}+02,7.32 \mathrm{E}+02,1.93 \mathrm{E}+03$, 3.21E+02, 7.40E+01,
raper weighted geometric mean of the VPA populations:
$2.37 \mathrm{E}+04,1.94 \mathrm{E}+04,1.63 \mathrm{E}+04,1.27 \mathrm{E}+04,8.58 \mathrm{E}+03,5.26 \mathrm{E}+03,3.16 \mathrm{E}+03,1.74 \mathrm{E}+03$, 8. $64 \mathrm{E}+02,4.18 \mathrm{E}+02$,

Standard error of the weighted Log(VPA populations) :


Log catchability residuals.

Fleet : SUMMER SURVEY

| Age | , | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002, | 2003, | 2004, | 2005 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  | No dat | for this fleet at this age |  |  |  |  |  |  |  |  |
| 1 | , | 1.22, | . 27 , | -. 14, | -. 20, | .13, | .17, | . 44 , | . 21, | -. 26 , | 32 |
| 2 | , | .16, | . 65, | . 05, | -. 16, | . 25 , | . 30, | . 20 , | .19, | . 52, | 23 |
| 3 | , | . 34, | .18, | -. 40, | 1.52, | . 20, | . 40 , | . 36, | -. 15, | -. 22 , | 04 |
| 4 | , | -. 45, | . 42 , | . 03, | -. 53, | -. 71, | . 26, | .11, | . 33, | -. 18, | 17 |
| 5 | , | -. 21 , | -.06, | . 02, | . 07 , | -. 20 , | -1.01, | . 09 , | . 51 , | . 23, | 00 |
| 6 | , | .19, | . 41, | -. 29, | . 06 , | . 09 , | -.35, | -.53, | -.15, | -. 10, | 74 |
| 7 | , | -.04, | -.37, | . 95 , | . 28 , | . 05 , | . 00 , | -. 36, | -. 30 , | -. 45, | 24 |
| 8 | , | -.09, | . 14, | . 61, | . 43, | . 29, | -.08, | -. 26 , | . 40 , | -. 75 , | -1.21 |
| Age | , | 2006, | 2007, | 2008, | 2009, | 2010, | 2011, | 2012, | 2013, | 2014, | 2015 |
| 0 |  | No dat | for | fle | at | s age |  |  |  |  |  |
| 1 |  | -.21, | . 05 , | . 38 , | . 57, | . 41, | -2.28, | -.49, | -. 32, | -. 06 , | -. 21 |
| 2 | , | . 57, | 1.16, | . 03, | -. 17, | . 27, | . 22 , | -2.09, | -2.20, | -. 44 , | . 27 |
| 3 | , | -.65, | -. 61, | -. 16, | -.99, | . 39, | . 12, | . 10, | -.35, | -. 22 , | . 11 |
| 4 |  | -. 02, | -. 66, | . 22, | . 36, | . 47, | -. 27 , | . 19, | . 32, | -. 20 , | . 15 |
| 5 |  | . 05 , | -. 27 , | -. 71, | . 06 , | . 06 , | -. 05 , | -.36, | . 70 , | . 28 , | . 78 |
| 6 |  | . 26 , | .15, | . 00 , | -. 28, | . 29 , | -. 31, | -. 41, | . 25 , | . 24, | -. 26 |
| 7 |  | . 32, | . 00 , | . 28, | .19, | . 09, | -. 32, | -. 05, | -.14, | -. 10, | -. 62 |
| 8 |  | -. 50, | -. 68, | . 20 , | -. 15, | .19, | -. 17, | -. 60, | . 79, | . 07 , | . 21 |

## Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age | 1, | 2, | 3 , | 4, | 5, | 6, | 7, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 8 |  |  |  |  |  |  |  |
| Mean Log q, | -5.0345, | -5.4842, | -5.6978, | -5.6563, | -5.7041, | -5.7964, | -5.7964, |
| 5.7964, |  |  |  |  |  |  |  |
| S.E(Log q) , | .6641, | . 8064 , | . 5220, | . 3637 , | . 4181, | . 3231 , | . 3521 , |
| . 5034, |  |  |  |  |  |  |  |

Table 5.8Faroe haddock 2016 xsa (cont.)

Regression statistics :
Ages with $q$ independent of year class strength and constant w.r.t. time
Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Q


Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 0, | 1, | 2, | 3, | 4, | 5, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -6.0219, | -5.3341, | -5.8327, | -5.8721, | -6.0498, | -6.2134, |
| S.E (Log q), | .8089, | .5742, | .6749, | .4384, | .7231, | .6814, |

Table 5.8 Faroe haddock 2016 xsa (cont.)

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 0, | .89, | .804, | 6.42, | .73, | 23, | .73, | -6.02, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1, | 1.15, | -1.247, | 4.71, | .77, | 23, | .65, | -5.33, |
| 2, | .93, | .637, | 6.08, | .79, | 23, | .64, | -5.83, |
| 3, | 1.03, | -.373, | 5.78, | .89, | 23, | .46, | -5.87, |
| 4, | .95, | .451, | 6.18, | .79, | 23, | .70, | -6.05, |
| 5, | 1.02, | -.181, | 6.17, | .78, | 23, | .71, | -6.21, |
| 6, | .99, | .071, | 6.45, | .75, | 23, | .76, | -6.44, |

Terminal year survivor and $F$ summaries :
Age 0 Catchability constant w.r.t. time and dependent on age
Year class $=2015$


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | S.e, | s.e, | , | Ratio, |  |
| $51118 .$, | .83, | .00, | 1, | .000, | .000 |

Age 1 Catchability constant w.r.t. time and dependent on age
Year class $=2014$

| Fleet, | Estimated, Survivors, | $\begin{aligned} & \text { Int, } \\ & \text { s.e, } \end{aligned}$ | $\begin{aligned} & \text { Ext, } \\ & \text { s.e, } \end{aligned}$ | Var, <br> Ratio, | N, | Scaled, Weights, | Estimate F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SUMMER SURVEY | 5536., | . 680, | . 000, | . 00 , | 1, | . 331, | . 000 |
| SPRING SURVEY SHIFTE, | 7573., | . 478 , | . 268 , | . 56 , | 2, | . 669, | . 000 |
| F shrinkage mean | $0 .$, | . 50, |  |  |  | . 000, | . 000 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $6828 .$, | .39, | .19, | 3, | .478, | .000 |

## Table 5.8Faroe haddock 2016 xsa (cont.)

| Year class $=2013$ |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fleet, | Estimated, | Int, | Ext, | Var, | N, | Scaled, | Estimated |
| , | Survivors, | s.e, | s.e, | Ratio, | , | Weights, | F |
| SUMMER SURVEY | 8543., | .525, | .159, | . 30, | 2, | .254, | . 040 |
| SPRING SURVEY SHIFTE, | 7056., | . 393 , | .175, | . 44 , | 3, | .454, | . 048 |
| F shrinkage mean , | 8957., | . 50,1, |  |  |  | . 292 , | . 038 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors, Int, | , Ext, | N, Var, | F |  |  |  |  |
| at end of year, s.e, | , s.e, | , Ratio, |  |  |  |  |  |
| 7942., .27, | , .10, | 6, .370, | . 043 |  |  |  |  |
| Age 3 Catchability | constant w | . time and | depend | nt on ag |  |  |  |
| Year class $=2012$ |  |  |  |  |  |  |  |
| Fleet, | Estimated, | Int, | Ext, | Var, | N, | Scaled, | Estimated |
| , | Survivors, | s.e, | s.e, | Ratio, | , | Weights, | F |
| SUMMER SURVEY , | 4106., | . 375 , | .172, | . 46 , | 3, | .303, | . 225 |
| SPRING SURVEY SHIFTE, | 5441., | .295, | .157, | . 53, | 4, | .486, | . 174 |
| F shrinkage mean , | 4021., | .50, , , , |  |  |  | .211, | . 229 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors, Int, | , Ext, | N, Var, | F |  |  |  |  |
| at end of year, s.e, | s.e, | , Ratio, |  |  |  |  |  |
| 4688., .21, | , .10, | 8, .496, | . 200 |  |  |  |  |
| Age 4 Catchability c | constant w.r. | time and d | ependen | t on age |  |  |  |
| Year class $=2011$ |  |  |  |  |  |  |  |
| Fleet, | Estimated, | Int, | Ext, | Var, | N, | Scaled, | Estimated |
| , | Survivors, | s.e, | s.e, | Ratio, | , | Weights, | F |
| SUMMER SURVEY | 701., | . 266 , | .380, | 1.43, | 4, | .442, | . 344 |
| SPRING SURVEY SHIFTE, | 1259., | . 275 , | . 273 , | .99, | 5, | . 374 , | . 206 |
| F shrinkage mean , | 717., | . 50, , , , |  |  |  | .184, | . 337 |


| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| 877., | .18, | .21, | 10, | 1.138, | .284 |

## Table 5.8Faroe haddock 2016 xsa (cont.)



Table 5.8Faroe haddock 2016 xsa (cont.)


Table 5.9 Faroe haddock. Fishing mortality (F) at age.




| . 0000 , | $\begin{aligned} & 0, \\ & .0000, \end{aligned}$ | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| . 0006 , | $\begin{aligned} & 1, \\ & .0000, \end{aligned}$ | . 0014, | . 0000 , | . 0000 , | . 0002 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , |
| . 0329, | $\begin{aligned} & 2, \\ & .0280, \end{aligned}$ | . 0908, | . 0108, | . 0010, | . 0004 , | . 0325 , | . 0237, | . 0383, | . 0252 , |
| . 1167, | $\begin{aligned} & 3, \\ & .1695, \end{aligned}$ | . 1878, | . 1128, | . 0547 , | . 0458 , | . 0285 , | . 1374 , | . 4618, | . 1917, |
| . 3896 , | $\begin{aligned} & 4, \\ & .2392 \end{aligned}$ | . 3810 , | . 1815 , | . 1665 , | . 1255 , | . 2025 , | . 1314 , | . 3709, | . 3481 , |
|  | 5, | . 2216 , | . 5273, | . 2116 , | .1913, | . 2750 , | . 2112, | . 2918, | . 3498 , |
| . 2171 , | $\begin{aligned} & .3475, \\ & 6, \end{aligned}$ | . 2871 , | . 7246 , | . 3820 , | . 1409 , | . 2136, | . 2264 , | . 2775 , | . 1383, |
| . 3336 , | $\begin{aligned} & .4163, \\ & 7, \end{aligned}$ | . 1601 , | . 3904 , | . 5760, | . 2721 , | . 1702, | . 2004 , | . 2524 , | . 2991 , |
| . 0853 , | $\begin{aligned} & .2084, \\ & 8, \end{aligned}$ | . 2539, | . 3788, | . 4969 , | . 3303, | . 3954 , | . 0920, | . 2266 , | . 3102 , |
| . 2929 , | $\begin{aligned} & .1720, \\ & 9, \end{aligned}$ | . 2621 , | . 4437 , | . 3690 , | . 2130 , | . 2526 , | . 1730 , | . 2854 , | . 2907 , |
| . 2651 , | .2782, $+g p$, | . 2621 , | . 4437 , | . 3690 , | . 2130 , | . 2526 , | . 1730 , | . 2854 , | . 2907 , |
| . 2651 , | . 2782 , |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { FBAR } \\ & .2285, \end{aligned}$ | $\begin{aligned} & 3-7, \\ & .2762, \end{aligned}$ | . 2476 , | . 3873, | . 2782 , | . 1551 , | . 1780, | . 1814 , | . 3309, | . 2654 , |

Table $5.9 \quad$ Faroe haddock. Fishing mortality (F) at age (cont.).

| 1994, | $\begin{gathered} \text { Table } 8 \\ \text { YEAR, } \\ 1995, \end{gathered}$ | $\begin{gathered} \text { Fishing } \\ \text { 1986, } \end{gathered}$ | $\begin{gathered} \text { mortality } \\ 1987, \end{gathered}$ | (F) at 1988, | 1989, | 1990, | 1991, | 1992, | 1993, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |
|  | 0 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , |
| . 0000 , | . 0000 , |  |  |  |  |  |  |  |  |
|  | 1 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0061 , |
| . 0000 , | $\begin{aligned} & .0000, \\ & 2, \end{aligned}$ | . 0097 , | . 0337 , | . 0394 , | . 0050 , | . 0125, | . 0290 , | . 0167, | .0709, |
| . 0489 , | $\begin{aligned} & .0094, \\ & 3, \end{aligned}$ | . 0941 , | . 0927 , | . 0680 , | .1207, | . 1313, | . 1654 , | . 0745 , | .1662, |
| . 1645 , | $\begin{aligned} & .1053, \\ & 4, \end{aligned}$ | . 2491 , | . 1846 , | .1863, | .1363, | . 2211, | . 2724 , | . 1782 , | .1842, |
| . 2582 , | $\begin{aligned} & .3128, \\ & 5, \end{aligned}$ | . 2598 , | . 2623 , | . 2368 , | . 3328, | . 2335 , | . 2185, | . 2763 , | .1865, |
| . 1484 , | $\begin{aligned} & .3077, \\ & 6, \end{aligned}$ | . 3589 , | . 3082 , | . 3062 , | . 3210, | . 3575 , | . 3178, | . 2604 , | . 2067 , |
| . 2122 , | $\begin{aligned} & .1846, \\ & 7, \end{aligned}$ | . 1573, | . 4748 , | . 2083, | . 5175, | . 4240 , | . 4045 , | . 2681 , | .1979, |
| . 2518, | $\begin{aligned} & .2248, \\ & 8, \end{aligned}$ | . 5179, | . 5848 , | . 2381 , | . 3887 , | . 4633, | . 2690 , | . 2310, | .1588, |
| . 2433, | $\begin{aligned} & .2712, \\ & 9, \end{aligned}$ | . 3104 , | . 3653 , | . 2363 , | . 3414, | . 3420 , | . 2981, | . 2440 , | .1876, |
| .2239, | $\begin{aligned} & .2616, \\ & +g p, \end{aligned}$ | . 3104 , | . 3653 , | . 2363 , | . 3414, | . 3420 , | . 2981, | . 2440 , | .1876, |
| $\begin{gathered} .2239, \\ \text { FBAR } \\ .2070, \end{gathered}$ | $\begin{aligned} & .2616, \\ & 3-7, \\ & .2271, \end{aligned}$ | . 2239, | . 2645 , | . 2011, | . 2857 , | . 2735, | . 2757 , | . 2115, | .1883, |



| . 3250, | $\begin{aligned} & 4, \\ & .2839, \end{aligned}$ | . 1844 , | . 1238 , | . 2914 , | . 2786 , | . 3974, | . 3871, | . 3674 , | .1999, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| . 3227 , | $\begin{aligned} & 5, \\ & .5263, \end{aligned}$ | . 2866 , | . 3057 , | . 1333, | . 2875, | . 3179, | . 3244 , | . 3257 , | . 2633 , |
| . 1880, | $\begin{aligned} & 6, \\ & .1408, \end{aligned}$ | . 5413, | . 4150, | . 2681 , | . 2372 , | . 4381, | . 3634 , | . 3140, | . 3374, |
| . 3779, | $\begin{aligned} & 7, \\ & .1294, \end{aligned}$ | .6995, | .5909, | . 3915 , | . 3415, | . 4334, | .4919, | . 4369, | . 3686 , |
| . 2527 , | $\begin{aligned} & 8, \\ & .2853, \end{aligned}$ | . 3657 , | . 5473, | . 5931, | . 2924 , | . 3784 , | .5499, | . 3018, | . 6516, |
| . 2123, | $\begin{aligned} & 9 \\ & .2909, \end{aligned}$ | .6209, | . 4352 , | . 2943, | . 2545 , | . 2867 , | . 3385, | . 5687, | . 3931 , |
|  | +gp, | .6209, | . 4352 , | . 2943, | . 2545 , | . 2867 , | . 3385, | . 5687, | . 3931 , |
| $\begin{array}{r} .2123, \\ \text { FBAR } \end{array}$ | .2909, $3-7$, | . 3573, | . 3271 , | . 2361 , | . 2666 , | . 3790, | . 3777, | . 3128, | . 2670 , |

Table $5.10 \quad$ Faroe haddock. Stock number (N) at age.

| Run title : FAROE HADDOCK (ICES |  |  |  | DIVISION | 5.b) |  | HAD_IND |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| At | 29/04/2016 | 11:08 |  |  |  |  |  |  |  |
|  |  | Terminal Fs derived using XSA (With F shrinkage) |  |  |  |  |  |  |  |
|  | Table 10 | Stock n | umber-at- | ge (start | of year) |  | N | mbers*10 |  |
|  | YEAR, | 1957, | 1958, | 1959, | 1960, | 1961, | 1962, | 1963, | 1964, |
|  | AGE |  |  |  |  |  |  |  |  |
|  | 0 , | 64927, | 54061, | 77651, | 58761, | 71715, | 45399, | 33843, | 30192, |
| 37948, |  |  |  |  |  |  |  |  |  |
|  | 1, | 47944, | 53158, | 44261, | 63576, | 48109, | 58715, | 37170, | 27709, |
| 24719, |  |  |  |  |  |  |  |  |  |
|  | 2 , | 35106, | 39212, | 43417, | 35763, | 51279, | 38537 , | 47362, | 30110, |
| 22644, |  |  |  |  |  |  |  |  |  |
|  | 3 , | 25440, | 25003, | 26445, | 31954, | 23796, | 34806, | 22837, | 26515, |
| 22585, |  |  |  |  |  |  |  |  |  |
|  | 4, | 20280, | 14377, | 13213, | 14717, | 16517, | 12850, | 15850, | 10638, |
| 14961, |  |  |  |  |  |  |  |  |  |
|  | 5, | 5517, | 8965, | 6632, | 6706, | 6028, | 8877, | 5786, | 6278, |
| 5182, |  |  |  |  |  |  |  |  |  |
|  | 6, | 2786, | 3055, | 4284, | 3570, | 3245, | 3182, | 5132, | 2708, |
| 3005, 7 |  |  |  |  |  |  |  |  |  |
|  | 7, | 1377, | 1472, | 1326, | 1839, | 1512, | 1476, | 1332, | 2809, |
| 1204, |  |  |  |  |  |  |  |  |  |
|  | 8, | 585, | 598, | 466, | 433, | 448, | 480, | 423, | 313, |
| 1641, |  |  |  |  |  |  |  |  |  |
|  | 9, | 252, | 274, | 224, | 168, | 135, | 153, | 148, | 114, |
| 77, |  |  |  |  |  |  |  |  |  |
|  | +gp, | 154, | 227, | 106, | 54, | 29, | 46 , | 45, | 16, |
| 14, |  |  |  |  |  |  |  |  |  |
|  | TOTAL, | 204367, | 200401, | 218024, | 217540, | 222811, | 204522, | 169929, | 137402, |
| 133981, |  |  |  |  |  |  |  |  |  |


|  | Table 10 | Stock number-at-age (start of year) |  |  |  |  | Numbers*10**-3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1974, | $\begin{aligned} & \text { YEAR, } \\ & 1975, \end{aligned}$ | 1966, | 1967, | 1968, | 1969, | 1970, | 1971, | 1972, | 1973, |
|  | AGE |  |  |  |  |  |  |  |  |
|  | 0 , | 81923, | 47768, | 53237, | 23136, | 49621, | 35418, | 78970, | 104847 |
| 83625, | 39127, |  |  |  |  |  |  |  |  |
|  | 1, | 31069, | 67073, | 39109, | 43587, | 18942, | 40627, | 28998, | 64655 |
| 85842, | 68466, |  |  |  |  |  |  |  |  |
|  | 2, | 20203, | 25356, | 54851, | 31975, | 35600, | 15457, | 33213, | 23702 |
| 52333, | 70052, |  |  |  |  |  |  |  |  |
|  | 3. | 17302, | 15563, | 19470, | 39587 , | 24022, | 27583, | 12006, | 26513 |
| 16410, | 37750, |  |  |  |  |  |  |  |  |
|  | 4, | 14613, | 11176, | 10566, | 12234, | 25590, | 15275, | 18608, | 6442 |
| 14092, | 10812, |  |  |  |  |  |  |  |  |
|  | 5, | 7604, | 7617 , | 6798, | 6106, | 5884, | 14996, | 8229, | 11454 |
| 4152, | 7946, |  |  |  |  |  |  |  |  |
|  | 6 , | 2937, | 3774 , | 4622, | 4187, | 3583, | 3348, | 9322, | 4288 |
| 6849, | 2992, |  |  |  |  |  |  |  |  |
|  | 7 , | 1366, | 1398, | 1800, | 2403, | 2084, | 1682, | 1572, | 6572 |
| 2680, | 4724, |  |  |  |  |  |  |  |  |
|  | 8 , | 377 , | 449, | 574, | 638, | 860, | 712, | 595, | 657 |
| 4427, | 1772, |  |  |  |  |  |  |  |  |
|  | 9. | 127, | 146, | 189, | 262, | 180, | 409, | 382, | 325 |
| 402, | 3141, |  |  |  |  |  |  |  |  |
|  | +gp, | 21, | 36, | 33, | 45, | 26, | 281, | 319, | 52 |
| 865,271678 | $1396,$ |  |  |  |  |  |  |  |  |
|  | TOTAL, | 177542, | 180355, | 191249, | 164160, | 166393, | 155787, | 192213, | 249509, |
|  | 248178, |  | 180355, |  |  |  |  |  |  |

Terminal Fs derived using XSA (With F shrinkage)
YEAR, 1976, 1977, 1978, 1979, 1980, 1981, 1982, 1983,

| 39475, | $\begin{aligned} & 0, \\ & 14060, \end{aligned}$ | 52360, | 4153, | 7376, | 5208, | 23620, | 29255, | 60791, | 58809, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 48148, | $\begin{aligned} & 1, \\ & 32319, \end{aligned}$ | 32035, | 42868, | 3400, | 6039, | 4264, | 19339, | 23952, | 49772, |
| 40750, | $\begin{aligned} & 2, \\ & 39398, \end{aligned}$ | 55970, | 26192, | 35098, | 2784, | 4944, | 3491, | 15833, | 19610, |
| 15656, | $\begin{aligned} & 3, \\ & 32282, \end{aligned}$ | 50715, | 41847, | 21213, | 28707, | 2278, | 3918, | 2791, | 12475, |
| 8432, | $\begin{gathered} 4, \\ 11406, \end{gathered}$ | 23712, | 34412, | 30607 , | 16443, | 22452, | 1813, | 2796, | 1440, |
| 832, | 5, | 6955, | 13262, | 23498, | 21215, | 11874, | 15012, | 1301, | 1580, |
|  | 6 , | 5265, | 4562, | 6408, | 15570, | 14345, | 7385, | 9951, | 796, |
| 912, | $\begin{aligned} & 548, \\ & 7, \end{aligned}$ | 2226, | 3235, | 1810, | 3581, | 11073, | 9486, | 4821, | 6173, |
| 567, | $\begin{aligned} & 535, \\ & 8, \end{aligned}$ | 3549, | 1553, | 1792, | 833, | 2233, | 7647, | 6356, | 3067, |
| 3747, | 927, | 1237, | 2254, | 870, | 893, | 490, | 1231, | 5711, | 4149, |
| 1841, | 2289, |  |  |  |  |  |  |  |  |
| 4566, | $\begin{aligned} & +g p, \\ & 4400, \end{aligned}$ | 1515, | 2613, | 1109, | 424, | 423, | 249, | 946, | 3460, |
| 164927, | $\begin{aligned} & \text { TOTAL, } \\ & 142340, \end{aligned}$ | 235538, | 176951, | 133181, | 101696, | 97996, | 98825, | 135250, | 161330, |

Table $5.10 \quad$ Faroe haddock. Stock number (N) at age (cont.).



|  | Table 10 | Stock | mber-at | ge (st | of ye |  |  | umbers* | **-3 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015, | $\begin{aligned} & \text { YEAR, } \\ & 2016, \end{aligned}$ | 2006, | 2007, | 2008, | 2009, | 2010, | 2011, | 2012, | 2013, | 2014, |
|  | AGE |  |  |  |  |  |  |  |  |  |
| 62435, | $0,$ $0,$ | 3966, | 3432, | 6301, | 17390, | 6731, | 4097, | 13129, | 15105, | 10186, |
| 8340, | $\begin{gathered} 1, \\ 51118, \end{gathered}$ | 3769, | 3247, | 2810, | 5159, | 14238, | 5511, | 3354, | 10749, | 12367, |
| 10125, | $\begin{array}{r} 2, \\ 6828, \end{array}$ | 7585, | 3085, | 2659, | 2295, | 4224, | 11657, | 4512, | 2746, | 8801, |
| 6990, | $\begin{aligned} & 3, \\ & 7942, \end{aligned}$ | 6833, | 5987, | 2457, | 2117, | 1854, | 3106, | 9390, | 3687, | 2173, |
| 1422, | $\begin{aligned} & 4, \\ & 4688, \end{aligned}$ | 16836, | 5191, | 4013, | 1827, | 1436, | 1116, | 1844, | 6819, | 2557, |
| 1512, | 5, 877, | 17514, | 11463, | 3755, | 2455, | 1132, | 790, | 620, | 1045, | 4572, |
| 2710, | $\begin{aligned} & 6, \\ & 732, \end{aligned}$ | 14338, | 10766, | 6913, | 2690, | 1508, | 675, | 468, | 367, | 658, |
| 446, | $\begin{gathered} 7, \\ 1928, \end{gathered}$ | 7200, | 6832, | 5820, | 4329, | 1738, | 797, | 384, | 280, | 214, |
| 120, | 8, 321, | 491, | 2929, | 3098, | 3221, | 2519, | 922, | 399, | 203, | 158, |
| 101, | $\begin{aligned} & 9, \\ & 74, \end{aligned}$ | 151, | 279, | 1387, | 1402, | 1969, | 1413, | 436, | 241, | 87 , |
| 87 , | $\begin{gathered} \text { +gp, } \\ \text { 115, } \end{gathered}$ | 165, | 28, | 168, | 235, | 701, | 715, | 287, | 401, | 276, |
| 94289, | $\begin{aligned} & \text { TOTAL, } \\ & 74622, \end{aligned}$ | 78847, | 53237, | 39381, | 43120 , | 38048 , | 30797, | 34822, | 41644, | 42048, |

Table 5.11. Faroe haddock. Stock summary of the 2016 VPA.


Table 5.12. Management options table INPUT DATA descriptions.
Stock size

The stock in numbers 2016 is taken directly from the 2016 XSA. The yearclass 2015 at age 2 (in 2017) is estimated from the 2016 XSA age 1 applying a natural mortality of 0.2 in forward calculation of the number using the standard VPA equation. The yearclass 2016 at age 2 (in 2018) is estimated as the geomean of the numbers-at-age 2 since 2005.

| AGE | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: |
| 2 | 6828 | 41852 | 5930 |
| 3 | 7942 |  |  |
| 4 | 4688 |  |  |
| 5 | 877 |  |  |
| 6 | 732 |  |  |
| 7 | 1928 |  |  |
| 8 | 321 |  |  |
| 9 | 74 |  |  |
| $10+$ | 115 |  |  |

Numbers in thousands (predicted values rounded).
Proportion mature at age
The proportion mature at age in 2016 is estimated as the average of the observed data in 2015 and 2016. For 2017 and 2018, the average of 2014-2016 is used.

| AGE | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: |
| 2 | 0.18 | 0.18 | 0.18 |
| 3 | 0.89 | 0.87 | 0.87 |
| 4 | 1.00 | 1.00 | 1.00 |
| 5 | 1.00 | 1.00 | 1.00 |
| 6 | 1.00 | 1.00 | 1.00 |
| 7 | 1.00 | 1.00 | 1.00 |
| 8 | 1.00 | 1.00 | 1.00 |
| 9 | 1.00 | 1.00 | 1.00 |
| $10+$ | 1.00 | 1.00 | 1.00 |

Table 5.12. Management options table-INPUT DATA descriptions (cont.).

Catch\&Stock weights at age

Catch and stock weights at age for all ages and for each of the years 2016-2018 are simply the average of the estimated point-values for 2013-2015 not re-scaled to 2015 since most weights have been fluctuating without any trend during the last 3 years ( no model was available to predict future mean weights at age).

| AGE | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: |
| 2 | 0.552 | 0.552 | 0.552 |
| 3 | 0.874 | 0.874 | 0.874 |
| 4 | 1.122 | 1.122 | 1.122 |
| 5 | 1.486 | 1.486 | 1.486 |
| 6 | 1.711 | 1.711 | 1.711 |
| 7 | 2.031 | 2.031 | 2.031 |
| 8 | 1.920 | 1.920 | 1.920 |
| 9 | 1.977 | 1.977 | 1.977 |
| $10+$ | 2.070 | 2.070 | 2.070 |

## Exploitation pattern

The exploitation pattern 2016 is estimated like last year as the average fishing mortality matrix in the 3 preceding years (2013-2015) from the final VPA in 2016, without re-scaling to the terminal year (2015) since fishing mortalities have been fluctuating without any general trend during the last 3 years; the same exploitation pattern was used for all 3 years.

| AGE | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: |
| 2 | 0.0357 | 0.0357 | 0.0357 |
| 3 | 0.1965 | 0.1965 | 0.1965 |
| 4 | 0.2696 | 0.2696 | 0.2696 |
| 5 | 0.3708 | 0.3708 | 0.3708 |
| 6 | 0.2221 | 0.2221 | 0.2221 |
| 7 | 0.2920 | 0.2920 | 0.2920 |
| 8 | 0.3965 | 0.3965 | 0.3965 |
| 9 | 0.2988 | 0.2988 | 0.2988 |
| $10+$ | 0.2988 | 0.2988 | 0.2988 |

Table 5.13 Faroe haddock. Management option table - Input data

MFDP version 1
Run: jr1
Time and date: 15:31 20/04/2016
Fbar age range: 3-7

| 2016 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 6828 | 0.2 | 0.18 | 0 | 0 | 0.552 | 0.036 | 0.552 |
| 3 | 7942 | 0.2 | 0.89 | 0 | 0 | 0.874 | 0.197 | 0.874 |
| 4 | 4688 | 0.2 | 1 | 0 | 0 | 1.222 | 0.270 | 1.222 |
| 5 | 877 | 0.2 | 1 | 0 | 0 | 1.486 | 0.371 | 1.486 |
| 6 | 732 | 0.2 | 1 | 0 | 0 | 1.711 | 0.222 | 1.711 |
| 7 | 1928 | 0.2 | 1 | 0 | 0 | 2.031 | 0.292 | 2.031 |
| 8 | 321 | 0.2 | 1 | 0 | 0 | 1.920 | 0.397 | 1.920 |
| 9 | 74 | 0.2 | 1 | 0 | 0 | 1.977 | 0.299 | 1.977 |
| 10 | 115 | 0.2 | 1 | 0 | 0 | 2.070 | 0.299 | 2.070 |
| 2017 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 41852 | 0.2 | 0.18 | 0 | 0 | 0.552 | 0.036 | 0.552 |
| 3 |  | 0.2 | 0.87 | 0 | 0 | 0.874 | 0.197 | 0.874 |
| 4 |  | 0.2 | 1 | 0 | 0 | 1.222 | 0.270 | 1.222 |
| 5 |  | 0.2 | 1 | 0 | 0 | 1.486 | 0.371 | 1.486 |
| 6 |  | 0.2 | 1 | 0 | 0 | 1.711 | 0.222 | 1.711 |
| 7 |  | 0.2 | 1 | 0 | 0 | 2.031 | 0.292 | 2.031 |
| 8 |  | 0.2 | 1 | 0 | 0 | 1.920 | 0.397 | 1.920 |
| 9 |  | 0.2 | 1 | 0 | 0 | 1.977 | 0.299 | 1.977 |
| 10 |  | 0.2 | 1 | 0 | 0 | 2.070 | 0.299 | 2.070 |
| 2018 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 5930 | 0.2 | 0.18 | 0 | 0 | 0.552 | 0.036 | 0.552 |
| 3 |  | 0.2 | 0.87 | 0 | 0 | 0.874 | 0.197 | 0.874 |
| 4 |  | 0.2 | 1 | 0 | 0 | 1.222 | 0.270 | 1.222 |
| 5 |  | 0.2 | 1 | 0 | 0 | 1.486 | 0.371 | 1.486 |
| 6 |  | 0.2 | 1 | 0 | 0 | 1.711 | 0.222 | 1.711 |
| 7 |  | 0.2 | 1 | 0 | 0 | 2.031 | 0.292 | 2.031 |
| 8 |  | 0.2 | 1 | 0 | 0 | 1.920 | 0.397 | 1.920 |
| 9 |  | 0.2 | 1 | 0 | 0 | 1.977 | 0.299 | 1.977 |
| 10 |  | 0.2 | 1 | 0 | 0 | 2.070 | 0.299 | 2.070 |

Input units are thousands and kg - output in tonnes

Table $5.14 \quad$ Faroe haddock. Management option table - Results
MFDP version 1
Run: jr1
Index file 20/04/2016
Time and date: 15:31 20/04/2016
Fbar age range: 3-7

2016

| Biomass | SSB | FMult | FBar | Landings |
| ---: | ---: | ---: | ---: | ---: |
| 23910 | 20056 | 1 | 0.2702 | 4251 |


| 2017 <br> Biomass |  |  |  | 2018 |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 43375 | 23818 | FMult | FBar | Landings | Biomass | SSB |
|  | 23818 | 0.1 | 0 | 0 | 53195 | 46619 |
|  | 23818 | 0.2 | 0.054 | 597 | 52532 | 45970 |
|  | 23818 | 0.3 | 0.0811 | 1178 | 51887 | 45338 |
|  | 23818 | 0.4 | 0.1081 | 2297 | 51257 | 44723 |
|  | 23818 | 0.5 | 0.1351 | 2835 | 50043 | 44123 |
|  | 23818 | 0.6 | 0.1621 | 3359 | 49461 | 42968 |
|  | 23818 | 0.7 | 0.1891 | 3870 | 48892 | 42412 |
|  | 23818 | 0.8 | 0.2161 | 4368 | 48337 | 41871 |
|  | 23818 | 0.9 | 0.2432 | 4854 | 47795 | 41342 |
|  | 23818 | 1 | 0.2702 | 5328 | 47267 | 40827 |
|  | 23818 | 1.1 | 0.2972 | 5790 | 46751 | 40325 |
|  | 23818 | 1.2 | 0.3242 | 6240 | 46247 | 39835 |
|  | 23818 | 1.3 | 0.3512 | 6680 | 45756 | 39356 |
|  | 23818 | 1.4 | 0.3783 | 7109 | 45276 | 38890 |
|  | 23818 | 1.5 | 0.4053 | 7527 | 44807 | 38434 |
|  | 23818 | 1.6 | 0.4323 | 7936 | 44350 | 37990 |
|  | 23818 | 1.7 | 0.4593 | 8335 | 43903 | 37556 |
|  | 23818 | 1.8 | 0.4863 | 8724 | 43466 | 37132 |
|  | 23818 | 1.9 | 0.5134 | 9104 | 43039 | 36719 |
|  | 23818 | 2 | 0.5404 | 9476 | 42622 | 36315 |

Input units are thousands and kg - output in tonnes

Table 5.15
Faroe haddock. Long-term Prediction - Input data
MFYPR version 1
Run: jr2
Index file 20/04/2016
Time and date: 16:20 20/04/2016
Fbar age range: 3-7

| Age | M | Mat | PF | PM | SWt | Sel | CWt |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.2 | 0.064 | 0 | 0 | 0.569 | 0.036 | 0.569 |
| 3 | 0.2 | 0.527 | 0 | 0 | 0.809 | 0.197 | 0.809 |
| 4 | 0.2 | 0.926 | 0 | 0 | 1.075 | 0.270 | 1.075 |
| 5 | 0.2 | 0.993 | 0 | 0 | 1.376 | 0.371 | 1.376 |
| 6 | 0.2 | 0.999 | 0 | 0 | 1.650 | 0.222 | 1.650 |
| 7 | 0.2 | 1.000 | 0 | 0 | 1.913 | 0.292 | 1.913 |
| 8 | 0.2 | 1.000 | 0 | 0 | 2.117 | 0.397 | 2.117 |
| 9 | 0.2 | 1.000 | 0 | 0 | 2.333 | 0.299 | 2.333 |
| 10 | 0.2 | 1.000 | 0 | 0 | 2.629 | 0.299 | 2.629 |

Weights in kilograms

## Table 5.16

Faroe haddock. Long-term Prediction - Results
MFYPR version 1
Run: jr2
Time and date: 16:20 20/04/2016
Yield per results

| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 0 | 0 | 0 | 5.5167 | 8.2937 | 4.1398 | 7.3883 | 4.1398 | 7.3883 |
| 0.1 | 0.027 | 0.1035 | 0.1726 | 5.0011 | 7.0969 | 3.6269 | 6.1942 | 3.6269 | 6.1942 |
| 0.2 | 0.054 | 0.1831 | 0.2925 | 4.605 | 6.2018 | 3.2336 | 5.3018 | 3.2336 | 5.3018 |
| 0.3 | 0.0111 | 0.2463 | 0.378 | 4.2911 | 5.5115 | 2.9224 | 4.614 | 2.9224 | 4.614 |
| 0.4 | 0.1081 | 0.2976 | 0.4402 | 4.0362 | 4.9658 | 2.6701 | 4.0708 | 2.6701 | 4.0708 |
| 0.5 | 0.1351 | 0.3403 | 0.4863 | 3.8248 | 4.5256 | 2.4613 | 3.6331 | 2.4613 | 3.6331 |
| 0.6 | 0.1621 | 0.3763 | 0.5208 | 3.6467 | 4.1643 | 2.2857 | 3.2743 | 2.2857 | 3.2743 |
| 0.7 | 0.1891 | 0.4071 | 0.547 | 3.4943 | 3.8634 | 2.1359 | 2.9758 | 2.1359 | 2.9758 |
| 0.8 | 0.2161 | 0.4338 | 0.567 | 3.3625 | 3.6096 | 2.0065 | 2.7243 | 2.0065 | 2.7243 |
| 0.9 | 0.2432 | 0.4572 | 0.5824 | 3.2472 | 3.3931 | 1.8936 | 2.5101 | 1.8936 | 2.5101 |
| 1 | 0.2702 | 0.4779 | 0.5943 | 3.1454 | 3.2066 | 1.7942 | 2.3259 | 1.7942 | 2.3259 |
| 1.1 | 0.2972 | 0.4963 | 0.6035 | 3.0548 | 3.0445 | 1.7061 | 2.166 | 1.7061 | 2.166 |
| 1.2 | 0.3242 | 0.5129 | 0.6107 | 2.9737 | 2.9024 | 1.6272 | 2.0261 | 1.6272 | 2.0261 |
| 1.3 | 0.3512 | 0.5278 | 0.6162 | 2.9005 | 2.777 | 1.5564 | 1.9029 | 1.5564 | 1.9029 |
| 1.4 | 0.3783 | 0.5414 | 0.6205 | 2.8341 | 2.6656 | 1.4922 | 1.7937 | 1.4922 | 1.7937 |
| 1.5 | 0.4053 | 0.5538 | 0.6238 | 2.7736 | 2.5661 | 1.434 | 1.6962 | 1.434 | 1.6962 |
| 1.6 | 0.4323 | 0.5652 | 0.6263 | 2.7181 | 2.4767 | 1.3808 | 1.6089 | 1.3808 | 1.6089 |
| 1.7 | 0.4593 | 0.5757 | 0.6282 | 2.6672 | 2.3959 | 1.332 | 1.5302 | 1.332 | 1.5302 |
| 1.8 | 0.4863 | 0.5854 | 0.6296 | 2.62011 | 2.3227 | 1.287 | 1.4589 | 1.287 | 1.4589 |
| 1.9 | 0.5134 | 0.5944 | 0.6306 | 2.5765 | 2.2559 | 1.2456 | 1.3941 | 1.2456 | 1.3941 |
| 2 | 0.5404 | 0.6028 | 0.6312 | 2.5359 | 2.1949 | 1.2071 | 1.335 | 1.2071 | 1.335 |


| Reference point | F multiplier | Absolute $F$ |
| :--- | ---: | ---: |
| Fbar(3-7) | 1 | 0.2702 |
| FMax | 2.2356 | 0.604 |
| F0.1 | 0.7341 | 0.1984 |
| F35\%SPR | 0.8628 | 0.2331 |
| Flow | -99 |  |
| Fmed | 0.812 | 0.2194 |
| Fhigh | 3.0515 | 0.8245 |

Weights in kilograms

Table $5.17 \quad$ Haddock biomass (age 2+) 1914-2015 in tons. Year label is sum of first row and first column.

|  | 1900 | 1925 | 1950 | 1975 | 2000 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 0 |  | 51433 | 81427 | 138898 | 107887 |
| 1 |  | 62983 | 67514 | 143617 | 143907 |
| 2 |  | 61870 | 66118 | 121035 | 150576 |
| 3 |  | 49631 | 73973 | 120568 | 137270 |
| 4 |  | 41313 | 75493 | 99492 | 124158 |
| 5 |  | 38518 | 88337 | 87629 | 88241 |
| 6 |  | 38412 | 91960 | 78954 | 64448 |
| 7 |  | 41379 | 90264 | 68298 | 46441 |
| 8 |  | 32637 | 92975 | 63951 | 33351 |
| 9 |  | 36717 | 89969 | 100634 | 24503 |
| 10 |  | 42386 | 96422 | 93926 | 20905 |
| 11 |  | 48161 | 93296 | 98457 | 18478 |
| 12 |  | 63897 | 98262 | 87574 | 16375 |
| 13 |  | 63738 | 90204 | 77340 | 16814 |
| 14 | 68122 | 50969 | 75561 | 69437 | 18411 |
| 15 | 60228 | 67182 | 71884 | 53438 | 25984 |
| 16 | 59115 | 84454 | 68774 | 38613 |  |
| 17 | 58850 | 113753 | 77101 | 28972 |  |
| 18 | 145265 | 103316 | 87971 | 28639 |  |
| 19 | 195704 | 97594 | 94878 | 27298 |  |
| 20 | 92229 | 90918 | 92142 | 86938 |  |
| 21 | 101713 | 126946 | 92929 | 111696 |  |
| 22 | 63831 | 103104 | 91506 | 106263 |  |
| 23 | 61076 | 91342 | 98976 | 91219 |  |
| 24 | 44757 | 80374 | 116873 | 78812 |  |

### 5.18 Figures



Figure 5.1. Haddock in ICES Division 5.b. Landings by all nations 1904-2015. Horisontal line average for the whole period.


Figure 5.2. Faroe haddock. Cumulative Faroese landings from 5.b.


Figure 5.3. Faroe haddock. Contribution (\%) by fleet to the total Faroese landings 2015.

Faroe Haddock LN(catch at age in numbers) for YC's 1948 onwards


Figure 5.4.


Figure 5.5. Faroe haddock. Mean weight at age (2-7).

Faroe Haddock - Maturity at age 1982-2015


Figure 5.6. Faroe haddock. Maturity-at-age since 1982. Running 3-years average of survey observations.


Fig-
ure 5.7. Commercial cpue's for Pairtrawlers > $\mathbf{1 0 0 0} \mathbf{H P}$ and longliners $>100 \mathrm{HP}$.


Figure 5.8. Faroe haddock. cpue (kg/trawlhour) in the spring and summer surveys.



Figure 5.9. Distribution of Faroe haddock catches in the spring survey (page above) and in the summer survey (this page).


Figure 5.10. Faroe haddock. LN (c@age in numbers) in the spring survey.

## Faroe Haddock Summer Survey



Figure 5.11. Faroe haddock. LN (c@age in numbers) in the summer survey.

Faroe haddock. Spring survey log q residuals.


Faroe haddock. Summer survey log q residuals.


Figure 5.12. Faroe haddock survey $\log q$ residuals.


Figure 5.13. Faroe haddock. Retrospective analysis on the 2015 XSA.

### 6.1 Stock description and management units.

See the stock annex.

### 6.2 Scientific data

### 6.2.1 Trends in landings and fisheries

Nominal landings of saithe from Faroese grounds (Division 5.b) have varied cyclically between 10000 t and 68000 t since 1961. After a third high of about 60000 t in 1990, landings declined steadily to 20000 t in 1996. Since then landings have increased to 68 000 tonnes in 2005 (Table 6.2.1.1, Figure 6.2.1.1) but has declined to 57000 tonnes in 2008 and 2009. After a substantial drop in landings in 2011 which was the lowest observed since 1999 ( 33000 t ) landings increased by $20 \%$ in 2012 up to 35000 t . Since 2011 landings have remained below historical average ( 37000 t.) The total tonnage in 2015 is 25000 t .

Since the introduction of the 200 miles EEZ in 1977, the saithe fishery has been prosecuted mostly by Faroese vessels. The principal fleet consists of large pairtrawlers $(>1000 \mathrm{HP})$, which have a directed fishery for saithe, about $50-77 \%$ of the reported landings in 1992-2011 (Table 6.2.1.2). The smaller pairtrawlers ( $<1000 \mathrm{HP}$ ) and single trawlers ( $400-1000 \mathrm{HP}$ ) have a more mixed fishery and they have accounted for about $10-20 \%$ of the total landings of saithe in the 1997-2011 period while the percentage of total landings by large single trawlers ( $>1000 \mathrm{HP}$ ) has declined drastically to just $1 \%$. Historically the catch composition by the pairtrawler fleet has accounted for about $75 \%$ of the total tonnage for saithe but since 2007 it has increased gradually up to $93 \%$ in 2015 due mainly to the gear-shifting of single-trawlers to pair-trawling. The share of catches by the jigger fleet was about $8 \%$ in the 1985-1998 period but has decreased to less than $0.5 \%$ since 2000 and it now accounts for only $2 \%$ of the total domestic landings for saithe in 2014. Foreign catches that have been reported to the Faroese Authorities but not officially reported to ICES are also included in the Working Group estimates. Catches in Subdivision IIa, which lies immediately north of the Faroes, have also been included. Little or no discarding is thought to occur in this fishery. Effort (measured as the ratio of nominal to used fishing days by the pair-trawl fleet segment) has diminished considerably in recent years. In the 2013/2014 fishing year only 58\% and $41 \%$ of fishing days were utilized in the inner and outer areas respectively while in the $2014 / 2015$ fishing year these ratios went up to $97 \%$ and $74 \%$, i.e. $29 \%$ of fishing days were not used in the last finished year season.
Cumulative landings of saithe for the domestic fleets since 2000 are shown in Figure 6.2.1.2. The period from 2011 to 2015 are among the poorest in the time-series. The progression of landings in the first three months of 2016 is below monthly averages and suggests a poor fishing year.

### 6.2.2 Catch-at-age

Catch-at-age is based on length, weight and otoliths samples from Faroese landings of small and large single and pairtrawlers, and landing statistics by fleet provided by the Faroese Authorities. Catch-at-age is calculated for each fleet by four-month periods and the total is raised by the foreign catches. Minor adjustments were made to the catch-at-age matrix for 2014 due to revised final catch statistics (Tables 6.2.2.1 and
6.2.2.2). Most of the age-disaggregated catch matrix is comprised of catches of the pairtrawl fleet. Since 2010 catch numbers is mostly comprised of age groups 4 to 6 whereas in the period from 2005 to 2009 it is mainly composed of age groups 4 to 8 . Numbers of 3 and 6 -years old were higher in 2015 than in 2014. While catches of 4 and 5 -years old saithe decreased from 2323 and 3143 thous. to 2269 and 2577 thous. respectively in 2015. Numbers of age 3 (recruiting age) in 2015 (2135 thous.) is the largest since 2010 (2324 thous.)

The sampling program and sampling intensity in 2015 as well as the approach used in compiling catch numbers is the same as in preceding years. Sampling levels of catches in 2014 was $8.9 \%$ and it went up to $9.9 \%$ in 2015 (Table 6.2.2.3.) The average amount sampled per tonnes landed since 2000 is $6.2 \%$.

### 6.2.3 Weight at age

Mean weights at age have varied by a factor of about 2 during since 1961. Mean weights at age were generally high during the early 1980s and they subsequently decreased from the mid-1980s to the early 1990s (Table 6.2.3.1 and Figure 6.2.3.1). Mean weights increased again in the period 1992-96 but have shown a general decrease thereafter. With the exception of 3-years old saithe all age groups were showing signs of increasing size since 2006. In 2011 age classes 4 to 6 were close or at long-term average. From 2012 to 2014 weight was below average for age groups 3 to 7 . Age classes 7 and older are in 2015 above mean values whereas younger age groups (3-5) are lower than average. Mean weight of the 2012 year class (age 3 in 2015) is estimated at 0.932 kg . which is the lowest ever observed in the time-series. On the other hand weight for 9 -years old saithe ( 6.715 kg .) is the largest since 1985. Mean weights at age in the stock are assumed equal to those in the catch.

### 6.2.4 Maturity-at-age

Maturity-at-age data from the spring survey is available from 1983 onward (Steingrund, 2003.) Due to poor sampling in 1988 the proportion mature for that year was calculated as the average of the two adjacent years. At the 2012 working group a model using maturity-at-age from the Faroese groundfish spring survey was implemented to derive smoothed trends in maturity by age and year. The fitting was done locally and the smoothing level was chosen as a trade-off between retaining the trend in maturities and reducing the data noise. For 1962-1982 the average maturity of predicted ogives of the 1983-2011 period was used (Table 6.2.4.1 and Figure 6.2.4.1.) Maturity ogives were low from the early and mid-1990s up to 2001 where they began to rise considerably and are above historical average since 2012.

Faroe saithe begins to mature at 3 years old, approximately $20 \%$ are mature at age 4, $50 \%$ at 5 years old and $100 \%$ are mature at age 9 and onwards.

### 6.2.5 Indices of stock size

### 6.2.5.1 Surveys

There are two annual groundfish surveys conducted in Faroese waters. The spring survey series (FGFS1) are available since 1994, while the summer survey (FGFS2) was initiated in 1996. The design for both bottom-trawl surveys is depth stratified with randomized stations covering the Faroe Plateau area. The total number of stations in the summer and spring is 100 and 200 respectively. Effort is recorded in terms of
minutes towed approximately 60 min . Large proportion of saithe is caught in relatively few hauls and the interannual variability of these hauls is considerable.

Survey catch rates (kg per hour), length composition and age-disaggregated indices are presented in figures 6.2.5.1.1 to 6.2.5.1.5. Both surveys suggest low abundances of saithe in mid- and late 1990's and increasing numbers from 2001 to 2005 although they differ in the order of this magnitude. Since 2007 the indices show that the saithe stock is at low levels while there are indications of a slight upward trend since 2011. Both surveys agreed not only in the direction but also in the magnitude of this positive trend. The most recent estimate of the spring survey suggest a slightly decrease in stock biomass in 2016 but given the uncertainty associated with the index the point estimate ought to be taken with cautioagreement between survey indices and the commercial series used in the model tuning is good. Both survey at age numbers agreed in the lack of year classes present in the stock since 2007. The spring index suggests that the 2002 year class (age 3 in 2015) is quite strong, which is confirmed by very abundant individuals of the same year class in 2016. Year-class strength in the summer index also suggests that the 2012 year class is strong.

Given the extreme schooling behaviour of saithe the internal consistency in the spring survey measured by the correlation of numbers in the data matrix for the same year class is reasonably good, with $\mathrm{R}^{2}$ close to 0.85 for the best defined age groups and below $\mathrm{R}^{2}=0.3$ for other age classes (Figure 6.2.5.1.6). Internal consistency in the age-disaggregated fall survey is displayed in figure 6.2.5.1.7. In terms of internal consistency the spring survey outperforms the fall survey.

### 6.2.5.2 Commercial cpue

The cpue series that has been used in the assessment since 2000 was introduced in 1998 (ICES C.M. 1998/ACFM:19), and consists of saithe catch-at-age and effort in hours, referred to as the pairtrawler series. A GLM model and a survey spatial scaling factor is used to standardized the cpue series (Stock Annex B.4., Benchmark report, WKROUND 2010.) The benchmark working group regarded this novel approach to developing the commercial series as reasonable (Benchmark report, WKROUND 2010.) Predicted annual cpues derived from this approach suggest that stock abundance was low in the 1990s and increased subsequently in the 2000's and a sharp downward trend from 2006 to 2011. Since 2012 the predicted cpue has remained remarkably stable at approximately $384 \mathrm{~kg} /$ hour (Figure 6.2.5.1.1)
The correlation between predicted cpue and the spring and summer surveys is $\mathrm{R}^{2}=0.55$ and $R^{2}=0.70$ respectively. The agreement between the survey indices measured by their correlation is estimated at $\mathrm{R}^{2}=0.36$.

The age composition indicates that the pair-trawl fleet targets mostly age groups 4 to 6. (Figure 6.2.5.2.1) There is a good agreement between age-disaggregated indices in the commercial index and indices of the same year class one year later (Figure 6.2.5.2.2) as measured by $\mathrm{R}^{2}>0.35$ for all age classes.

### 6.2.5.3 Information from the fishing industry

No additional information beyond the landings from the commercial fleet was presented for incorporation in the assessment.

### 6.3 Methods

The assessment model adopted at the benchmark assessment in 2010 is described in the Stock annex (Section C) and in the benchmark report (WKROUND 2010.) The 2010 XSA was calibrated with the standardized pairtrawlers with catchability independent of stock size for all ages, catchability independent of age for ages $\geq 8$, the shrinkage of the $S E$ of the mean $=2.0$, and no time tapered weighting. The tunings series used are shown in Table 6.3.1. Commercial catch-at age data (ages 3-14+, years 1961-2013) were calibrated in the XSA model using the commercial pair-trawl fleet (ages 3-11, years 1995-2013). XSA model diagnostics of the spaly run is presented in Table 6.3.2. Patterns in log-catchability residuals from the XSA model are relatively random but with large positive blocks in 2006-2010 for 3 to 4 age classes (Figure 6.3.1.). Residuals from a separable statistical model predicting catch numbers-at-age and survey data and modelling catchability, selectivity over 4 distinct periods and including a stock-recruitment relationship are also presented (Figure 6.3.3)

### 6.4 Reference points

### 6.4.1 Biological reference points and MSY framework

In 2014 at the WKMSYREF2 workshop the EqSim simulation framework was used to explore candidates to FmsY. The work was presented at the NWWG meeting in 2014 and the results agree with the previous simulations (see above) in that estimates of $\mathrm{F}_{\mathrm{MS}}$ are in the range of $\mathrm{F}_{\text {ms }}=0.30$ and $\mathrm{F}_{\text {ms }} \mathrm{y}=0.34$ and not as the present level of $\mathrm{F}_{\text {мsy }}=0.28$. In the 2014 meeting ACOM adopted the EqSim framework and agreed to set $\mathrm{FmSY}=0.30$, which agrees with the estimation of Fmed $=0.31$. Below it is an excerpt from the WKMSYREF2 report:

The EqSim framework fits three stock-recruit functions (Ricker, Beverton-Holt and Hockey-stick) on the bootstrap samples of the stock and recruit pairs from which approximate joint distributions of the model parameters can be made. The result of this is projected forward for a range of F's values and the last 50 years are retained to calculate summaries. Each simulation is run independently from the distribution of model and parameters. Error is introduced within the simulations by randomly generating process error about the constant stock recruit fit, and by using historical variation in maturity, natural mortality, weight at age, etc.

In the EqSim simulations the Hockey-Stick stock-recruit function were used assuming assessment and autocorrelation errors. Figures 6.4.1.1 and 6.4.1.2 illustrate the results of these simulations which suggest that candidates for FMSY are FMSY $=0.34$ (median yield) and FMSY $=0.30$ ( F that gives the maximum mean yield in the long term) if autocorrelation and assessment errors are included in the simulation framework. If errors are ignored then estimates for FMSY are predicted to FMSY $=0.38$ (median yield), FMSY $=0.35$ (maximum mean yield). No Blim is defined for faroe saithe but for the purposes of the analysis a value of $\mathrm{Bim}=\mathrm{Bp}_{\mathrm{p}} / 1.4$ was set for the simulations. A more detailed information of the simulations are available under http://www.ices.dk/community/groups/Pages/WKMSYREF2.aspx A summary is given in the table below.

|  | F | SSB | CATCH | option |
| :--- | :---: | :---: | :---: | :---: |
| Flim | 0.34 | 87327.43 | 36479.8 | ass. Error |
| Flim | 0.37 | 79116.87 | 35447.45 | ass. Error |
| Flim | 0.46 | 38905.3 | 22023.28 | ass. Error |
| MSY:median | 0.34 | 88565.78 | 36665.24 | ass. Error |
| Maxmeanland | 0.30 | 101372.9 | 37109.88 | ass. Error |
| FCrash | 0.41 | 63312 | 31637.31 | ass. Error |
| FCras50 | 0.52 | 855.73 | 550.19 | ass. Error |
| Flim | 0.40 | 78435.72 | 38526.07 | No ass. Error |
| Flim | 0.42 | 73052.08 | 37660.27 | No ass. Error |
| Flim | 0.50 | 38910.57 | 24279.75 | No ass. Error |
| MSY:median | 0.38 | 82329.53 | 38694.43 | No ass. Error |
| Maxmeanland | 0.35 | 90688.34 | 39167.13 | No ass. Error |
| FCrash5 | 0.43 | 69750.99 | 37114.99 | No ass. Error |
| FCrash50 | 0.54 | 2847.53 | 1910.51 | No ass. Error |

MSY and revised precautionary reference points (Section 2. Demersal stocks in the Faroe Area, Subsection 2.1.7 Faroe saithe) for Faroe saithe are listed below:

|  | BIOLOGICAL REFERENCE POINTS | NWWG 2012 |
| :--- | :---: | :---: |$\quad$ NWWG2015

The Yield/R and SSB-R calculations with respect to reference fishing mortalities (Fmax, Fmed and F0.1) is presented in the table below. The SSB-R plot in relation to Fhigh, Fmed and Flow is shown in Figure 6.4.1.3.

|  | FISH MORT <br> AGES 4-8 | YIELD/R | SSB/R |
| :--- | :---: | :---: | :---: |
| Average last 3 years | 0.35 | 1.28 | 2.84 |
| $F_{\max }$ | 0.42 | 1.29 | 2.36 |
| F $_{0.1}$ | 0.15 | 1.15 | 6.10 |
| $F_{\text {med }}$ | 0.31 | 1.28 | 3.29 |

### 6.5 State of the stock

Recruitment in the 1980s was close to the historical average ( 32 millions). The strongest year class since 1986 was produced in the 1990s and the average for that decade was about 28 million (Figures 6.5.1-6.5.4. and Tables 6.5.1 to 6.5.3). The 1998 ( 88 millions) and 1999 ( 106 millions) are the largest observed in the time-series. Since 2006 estimated recruitment has remained at low levels compared with the exceptionally high recruitment pulses observed from 2001-2005. However the 2012 year class (numbers of age-3 saithe in 2015) is estimated at 63 million and therefore far above the historical average of 32 million. Nevertheless the most recent recruitment estimate is highly unreliable and it contradicts with the estimate from a separable statistical model, which predicts recruitment at $\mathrm{N}_{3}(2015)=36$ million and thus in line with the present low productivity period.

Relatively low Fs during the 1960s and recruitment above average in early-1970s caused an increase in SSB well above the historical average around the mid-1970s while landings peaked to almost 58000 t . in 1973. Increasing Fs since 1980 lead to a decrease in the spawning-stock biomass of saithe throughout the mid-1980s although recruitment of the 1983 year class rose to 662000 millions, i.e. double the average from 1961 to 2014. The historically low SSB persisted in 1992-1998 and this along with low Fs caused landings to steeply decline to around 20000 tonnes in 1996. The SSB increased since 1999 to above 128000 t in 2005 with the maturation of the 1995, 1996, 1997 and 1999 year classes and decreased to 93000 t in 2009. The 2015 spaly assessment indicates that the point estimator of SSB (2014) is approximately 77000 t . From 2005 to 2013 SSB has been declining sharply but it has increased again since 2013 above $B_{\text {trigger }}=55000 \mathrm{t}$. due to improving maturity ogives and growth. Figure 6.5.6 illustrates the numbers of mature fish in the stock forage-groups from 3 to 9 in 2006, 2013 and 2014. It is quite clear that there has been a substantial increase in the numbers of mature fish over the age groups 3 to 6 a phenomenon supported by increased maturity ogives in recent years The separable catch-at-age model predicts $\operatorname{SSB}(2015)=68000 \mathrm{t}$.

In 2015 average fishing mortality over age groups 4 to 8 ( $\mathrm{Fbar}^{\mathrm{b}}$ ) is estimated at $\mathrm{F}(2015)=0.25$, which is the lowest since $1980(\mathrm{~F}=0.21)$ and therefore below $\mathrm{F}_{\mathrm{ms}}=0.30$. On the other hand the statistical model framework suggests that $\mathrm{F}(2015)=0.32$ is higher than that of the spaly assessment. The assessment model suggests a drop in fishing mortality since 2013 reflecting the abrupt decline in landings since 2011. The relation between stock and recruitment is presented in figure 6.5.7.

### 6.6 Short-term forecast

### 6.6.1 Input data

Population numbers-at-age 3 for the base short-term prediction is calculated as the geometric mean of estimated recruitment strength from 2010-2014. Natural mortality is set to constant 0.2 . Weight-at-age for 3 -years old saithe is predicted by the year-class strength (number of 3 -years old in the stock) with a 3 year time-lag (Eq. 1) whereas weight for ages 4-8 is estimated by weight-at-age the previous year from the same year class (Eq. 2) Weight for ages $9-14+$ is an average of the most 3 recent years. Diagnostics and results of the model are shown in Figures 6.6.1.1 and 6.6.1.2. For older age groups ( 9 to $14+$ ) a 3-year average is used.
$\mathrm{W} 3, \mathrm{y}=\alpha \mathrm{N} 3, \mathrm{y}-3+\beta$
for $\mathrm{a}=3$ (Eq. 1)
$\mathrm{Wa}+1, \mathrm{y}+1=\alpha \mathrm{Wa}, \mathrm{y}+\beta$
for $4 \leq \mathrm{a} \leq 8 \quad$ (Eq. 2)
$\mathrm{Wa}, \mathrm{y}=(\mathrm{Wa}-3, \mathrm{y} \mathrm{Wa}-2, \mathrm{y} \mathrm{Wa}-1, y) / 3 \quad$ for $9 \leq \mathrm{a} \leq 14+\quad$ (Eq. 3)
Proportion mature for 2016-2018 is taken as the average of predicted maturity ogives from 2014 and 2016. The exploitation pattern used is a 3 year average rescaled to last year as specified in the stock annex.

Input data for the prediction with management options for the spaly scenario are presented in Table 6.6.1.1.

### 6.6.2 Projection of catch and biomass

Results from predictions with management option is presented in Table 6.6.2.1 and Figure 6.6.2.1.

At status quo $\mathrm{F}=0.25$ landings would increase to 32 kt . in 2016 and 40 kt . in 2017 while spawning-stock biomass is expected to around 97 kt . in 2016 and increase to 126 kt . tonnes in 2017. Landings in 2016 are predicted to rely on the 2010, 2011 and 2012 year classes ( $69 \%$ ) while in the SSB these year classes will contribute to around $62 \%$ of the spawning biomass in 2016 (Figure 6.6.2.2.)

### 6.7 Yield-per-recruit and medium term forecasts

No medium term projections were performed for Faroe saithe.

## Input data to yield-per-recruit

The input data to long-term prediction are shown in Table 6.7.1.1.
Mean weights-at-age for 1981-2013 were used for the long-term projection. Natural mortality is set to constant 0.2. Proportion mature-at-age is taken as the average from 1983-2014.

The exploitation pattern was set equal to the average of the last five years (as suggested from ACFM, 2004). Results from the yield-per-recruit analysis is shown in Figure 6.7.1.1.

### 6.8 Uncertainties in assessment and forecast

In 2015 the amount of catch sampled was $9.9 \%$, which is regarded as adequate.
The assessment of Faroe saithe is relatively uncertain due to lack of good tuning data although the internal consistency in the commercial fleets used to calibrate the XSA model is reasonable considering the highly schooling and widely migrating behaviour of the species. The retrospective pattern (Figure 6.8.1) reveals some of the assessment uncertainty. It shows periods of over- and underestimation in average fishing mortality and consequently under- and overestimation in spawning-stock biomass. Overand underestimation seem to occur in periods of poor and high abundances respectively. Various factors could explain this phenomenon, e.g. by changes in the vertical distribution of the stock or changes in the selection pattern that have been observed in recent years. The retrospective plots show very small revisions in SSB and F in 2014 and 2015.

With respect to recruitment the retrospective trend suggests an overestimation of incoming year classes. To avoid large year-to-year fluctuations in the spawning-stock biomass (also dependent on age structure) a locally fitting model was implemented in 2012 to reduce variability of maturities.

### 6.9 Comparison with previous assessment and forecast

The 2015 assessment predicted recruitment for 205 to around 27 million while the observed year class strength was 63 million (Table 6.9.1). Fishing mortality was overestimated from $\mathrm{F}=0.31$ to $\mathrm{F}=0.25$. The spawning-stock biomass was overestimated by around $5 \%$. Landings for 2015 were predicted at Land(2015) $=35 \mathrm{kt}$. while actual observed catches in that year reached $\operatorname{Land}(2015)=25 \mathrm{kt}$ an overestimation of $40 \%$. Landings and recruitment estimates from the statistical model were however closer to the actual measurements $\operatorname{Land}(2015)=26 \mathrm{kt}$. and recruitment $\operatorname{Rec}(2015)=36$ mill.than the spaly run.

### 6.10 Management plans and evaluations

No management plan exists for saithe in Division 5.b

### 6.11 Management considerations

Management consideration for saithe is under the general section for Faroese stocks.
In 2014 ACOM adopted $\mathrm{Fms}_{\mathrm{m}}=0.30\left(\mathrm{~F}_{\mathrm{pa}}=0.30\right)$ presented at the NWWG meeting for the same year and produced in the WKMSYREF2 workshop on reference points. Btrigger is set at Bloss $=55 \mathrm{kt}$. (Btrigger $=55 \mathrm{kt})$.

### 6.12 Ecosystem considerations

No evidence is available to indicate that the fishery is impacting the marine environment. A PhD. project was initiated in 2008, with the aim of investigate the role of environmental indicators in the dynamics of Faroe saithe. The results and conclusions of the PhD will be available to the working group in future meetings.

### 6.13 Regulations and their effects

It seems to be no relationship between number of fishing days and fishing mortality, probably because of large fluctuations in catchability. Area restriction is an alternative to reduce fishing mortality- and this is used to protect small saithe in Faroese area.

### 6.14 Changes in fishing technology and fishing patterns

See section 6.2.

### 6.15 Changes in the environment

According to existing literature the productivity of the ecosystem clearly affects both cod and haddock recruitment and growth (Gaard et al., 2002), a feature outlined in Steingrund and Gaard (2005). The primary production on the Faroe Shelf (< 130 m depth), over the period May through June, varied interannually by a factor of five, giving rise to low- or high-productive periods of $2-5$ years duration (Steingrund and Gaard, 2005). The productivity over the outer areas seems to be negatively correlated with the strength of the Subpolar Gyre (Hátún et al., 2005; Hátún et al., 2009; Steingrund et al., 2010), which may regulate the abundance of saithe in Faroese waters (Steingrund and Hátún, 2008). When comparing a gyre index (GI) to saithe in Faroese waters there was a marked positive relationship between annual variations in GI and the total biomass of saithe lagged 4 years (Figure 6.15.1.)
There is a negative relationship between mean weight-at-age and the stock size of saithe in Faroese waters. This could be due to simple density-dependence, where there
is a competition for limited food resources. Stomach content data show that the food of saithe is dominated by blue whiting, Norway pout, and krill, and the annual variations in the stomach fullness are mainly attributable to variations in the feeding on blue whiting. There seems to be no relationship between stomach fullness and weights-atage for saithe (í Homrum et al. WD 2009).

### 6.16 References

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### 6.17 Tables

Table 6.2.1.1. Faroe saithe (Division 5.b). Nominal catches (tonnes round weight) by countries 19882015 as officially reported to ICES.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 94 | - | 2 | - | - | - | - | - | - | - | - | - | - | - |
| Estonia | - | - | - | - | - | - | - | - | - | 16 | - | - | - | - |
| Faroe Islands | 44402 | 43624 | 59821 | 53321 | 35979 | 32719 | 32406 | 26918 | 19267 | 21721 | 25995 | 32439 |  | 49676 |
| France 3 | 313 | - | - | - | 120 | 75 | 19 | 10 | 12 | 9 | 17 | - | 273 | 934 |
| Germany | - | - | - | 32 | 5 | 2 | 1 | 41 | 3 | 5 | - | 100 | 230 | 667 |
| German <br> Dem.Rep. | - | 9 | - | - | - | - | - | - | - | - | - | - | - | - |
| German Fed. Rep. | 74 | 20 | 15 | - | - | - | - | - | - | - | - | - | - | 5 |
| Greenland | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Ireland | - | - | - | - | - | - | - | - | - | - | - | 0 | 0 | 0 |
| Netherlands | - | 22 | 67 | 65 | - | - | - | - | - |  | - | 160 | 72 | 60 |
| Norway | 52 | 51 | 46 | 103 | 85 | 32 | 156 | 10 | 16 | 67 | 53 | - | - | - |
| Portugal | - | - | - | - | - | - | - | - | - | - | - | - | 20 | 1 |
| UK (Eng. \& W.) | - | - | - | 5 | 74 | 279 | 151 | 21 | 53 | - | 19 | 67 | 32 | 80 |
| UK (Scotland) | 92 | 9 | 33 | 79 | 98 | 425 | 438 | 200 | 580 | 460 | 337 | 441 | 534 | 708 |
| USSR/Russia 2 | - | - | 30 | - | 12 | - | - | - | 18 | 28 | - | - | - | - |
| Total | 45027 | 43735 | 60014 | 53605 | 36373 | 33532 | 33171 | 27200 | 19949 | 22306 | 26065 | 33207 | 1161 | 52131 |
| Working Group estimate 45 | 45285 | 44477 | 61628 | 54858 | 36487 | 33543 | 33182 | 27209 | 20029 | 22306 | 26421 | 33207 | 39020 | 51786 |
| COUNTRY | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| Denmark | - | - | - | - | 34 | - | - | - | - | - | - | - | - | - |
| Estonia | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Faroe Islands | 55165 | 47933 | 48222 | 71496 | 70696 | 64552 | 61117 | 61889 | 46686 | 32056 | 38175 | 28609 | 25474 | 26796 |
| France | 607 | 370 | 147 | 123 | 315 | 108 | 97 | 68 | 46 | 135 | 40 | 31 | 0 | 122 |
| Germany | 422 | 281 | 186 | 1 | 49 | 3 | 3 | 0 |  |  |  |  |  |  |
| Greenland | 125 | - |  |  | 73 | 239 | 0 | 1 |  |  | 1 |  |  |  |
| Irland | - | - | - | - | - | - | - | - |  |  |  |  |  |  |
| Iceland | - | - | - | - | - | - | - | 148 | - |  |  |  |  |  |
| Netherlands | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 |  |  |  |  |  |
| Norway | 77 | 62 | 82 | 82 | 35 | 81 | 38 | 23 | 28 |  |  |  | 165 | 40 |
| Portugal | - | - | 5 | - | - | - | - | - |  |  |  |  |  |  |
| Russia | 10 | 32 | 71 | 210 | 104 | 159 | 38 | 44 | 3 |  |  | 1 |  |  |
| UK (E/W/NI) | 58 | 89 | 85 | 32 | 88 | 4 | - | - |  |  |  |  |  |  |
| UK (Scotland) | 540 | 610 | 748 | 4322 | 1011 | 408 | 400 | 685 |  |  |  |  |  |  |
| United <br> Kingdom | - | - | - | - | - | - | - | - | 706 | 19 |  | 1 | 340 | 204 |
| Total | 57004 | 49377 | 49546 | 76266 | 72405 | 65557 | 61693 | 62858 | 47469 | 32210 | 38216 | 28642 | 25979 | 27262 |
| Working Group estimate | 53546 | 46555 | 46355 | 67967 | 66902 | 60785 | 57044 | 57949 | 43885 | 29658 | 35314 | 26463 | 23885 | 25128 |

Table 6．2．1．2．Faroe saithe（Division 5．b）．Total Faroese landings（rightmost column）and the contri－ bution（\％）by each fleet category（1985－2015）．Averages for 1985－2015 are given at the bottom．

|  | $\begin{aligned} & \tilde{4} \\ & \text { ¿} \\ & \text { z } \\ & \text { zü } \end{aligned}$ |  |  | $\begin{aligned} & \text { 号 } \\ & \text { ㄹ } \end{aligned}$ | $\begin{aligned} & \text { ̛山己 } \\ & \underline{U} \end{aligned}$ |  |  |  |  |  |  | N |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0.2 | 0.1 | 0.1 | 0.0 | 2.6 | 6.6 | 33.7 | 28.2 | 28.2 | 0.1 | 0.2 | 0.2 | 42598 |
| 1986 | 0.3 | 0.2 | 0.1 | 0.1 | 3.6 | 2.8 | 27.3 | 27.5 | 36.5 | 0.1 | 0.7 | 0.9 | 40107 |
| 1987 | 0.7 | 0.1 | 0.3 | 0.4 | 5.6 | 4.1 | 20.4 | 22.8 | 44.2 | 0.1 | 1.1 | 0.0 | 39627 |
| 1988 | 0.4 | 0.3 | 0.1 | 0.3 | 6.5 | 6.8 | 20.8 | 19.6 | 43.6 | 0.1 | 1.3 | 0.1 | 43940 |
| 1989 | 0.9 | 0.1 | 0.3 | 0.2 | 9.3 | 5.4 | 17.7 | 23.5 | 41.1 | 0.1 | 1.3 | 0.0 | 43624 |
| 1990 | 0.6 | 0.2 | 0.2 | 0.2 | 7.4 | 3.9 | 19.6 | 24.0 | 42.8 | 0.2 | 0.9 | 0.0 | 59821 |
| 1991 | 0.6 | 0.1 | 0.1 | 0.6 | 9.8 | 1.3 | 13.9 | 26.5 | 46.2 | 0.1 | 0.8 | 0.0 | 53321 |
| 1992 | 0.4 | 0.4 | 0.0 | 0.0 | 10.5 | 0.5 | 7.1 | 24.4 | 55.6 | 0.1 | 1.0 | 0.0 | 35979 |
| 1993 | 0.6 | 0.2 | 0.1 | 0.0 | 9.3 | 0.6 | 6.5 | 21.4 | 60.6 | 0.1 | 0.7 | 0.0 | 32719 |
| 1994 | 0.4 | 0.4 | 0.1 | 0.0 | 12.6 | 1.1 | 6.8 | 18.5 | 59.1 | 0.2 | 0.7 | 0.0 | 32406 |
| 1995 | 0.2 | 0.1 | 0.4 | 0.0 | 9.6 | 0.9 | 9.9 | 17.7 | 60.9 | 0.3 | 0.0 | 0.0 | 26918 |
| 1996 | 0.0 | 0.0 | 0.1 | 0.0 | 9.2 | 1.2 | 6.8 | 23.7 | 58.6 | 0.2 | 0.0 | 0.0 | 19267 |
| 1997 | 0.0 | 0.1 | 0.1 | 0.0 | 8.9 | 2.5 | 10.7 | 17.8 | 58.9 | 0.4 | 0.4 | 0.0 | 21721 |
| 1998 | 0.1 | 0.4 | 0.1 | 0.0 | 8.1 | 2.8 | 13.8 | 16.5 | 57.6 | 0.3 | 0.4 | 0.0 | 25995 |
| 1999 | 0.0 | 0.1 | 0.1 | 0.0 | 5.7 | 1.2 | 12.6 | 18.5 | 60.0 | 0.2 | 1.6 | 0.0 | 32439 |
| 2000 | 0.1 | 0.1 | 0.2 | 0.0 | 3.7 | 0.3 | 15.0 | 17.5 | 62.3 | 0.1 | 0.7 | 0.0 | 39020 |
| 2001 | 0.1 | 0.1 | 0.1 | 0.0 | 2.8 | 0.3 | 20.2 | 16.5 | 58.8 | 0.2 | 0.8 | 0.1 | 51786 |
| 2002 | 0.1 | 0.2 | 0.1 | 0.0 | 1.6 | 0.1 | 26.5 | 10.5 | 60.8 | 0.1 | 0.0 | 0.0 | 53546 |
| 2003 | 0.0 | 0.0 | 1.9 | 0.0 | 0.9 | 0.4 | 17.4 | 14.7 | 64.7 | 0.1 | 0.0 | 0.0 | 46555 |
| 2004 | 0.1 | 0.2 | 3.7 | 0.0 | 1.9 | 0.4 | 15.1 | 14.4 | 63.8 | 0.2 | 0.0 | 0.0 | 44605 |
| 2005 | 0.2 | 0.1 | 4.4 | 0.0 | 2.4 | 0.2 | 12.7 | 20.6 | 59.2 | 0.2 | 0.0 | 0.0 | 66394 |
| 2006 | 0.2 | 0.4 | 0.3 | 0.0 | 3.9 | 0.1 | 19.8 | 20.6 | 54.1 | 0.6 | 0.0 | 0.0 | 65394 |
| 2007 | 0.2 | 0.2 | 0.2 | 0.0 | 2.0 | 0.1 | 30.4 | 16.0 | 50.6 | 0.3 | 0.0 | 0.0 | 41341 |
| 2008 | 0.2 | 0.3 | 1.5 | 0.0 | 3.2 | 0.2 | 20.4 | 16.0 | 57.7 | 0.5 | 0.0 | 0.0 | 27475 |
| 2009 | 0.4 | 0.2 | 3.3 | 0.0 | 4.3 | 0.1 | 9.6 | 15.1 | 66.8 | 0.2 | 0.0 | 0.0 | 47122 |
| 2010 | 0.1 | 0.1 | 1.2 | 0.0 | 3.9 | 2.4 | 8.3 | 15.1 | 68.3 | 0.6 | 0.0 | 0.0 | 38293 |
| 2011 | 0.1 | 0.1 | 0.5 | 0.0 | 3.6 | 1.3 | 2.6 | 14.1 | 77.1 | 0.5 | 0.0 | 0.0 | 26854 |
| 2012 | 0.2 | 0.1 | 1.9 | 0.0 | 2.4 | 0.1 | 2.2 | 18.6 | 73.5 | 1.0 | 0.0 | 0.0 | 31633 |
| 2013 | 0.1 | 0.3 | 1.0 | 0.0 | 3.2 | 0.2 | 0.6 | 24.9 | 69.0 | 0.5 | 0.0 | 0.1 | 22339 |
| 2014 | 0.2 | 0.3 | 0.5 | 0.0 | 1.9 | 0.2 | 0.2 | 15.6 | 80.7 | 0.3 | 0.0 | 0.1 | 20793 |
| 2015 | 0.2 | 0.4 | 1.1 | 0.0 | 2.3 | 0.0 | 0.2 | 18.0 | 75.5 | 0.3 | 0.0 | 0.0 | 20956 |
| Avg． | 0.3 | 0.2 | 0.8 | 0.1 | 5.2 | 1.5 | 13.9 | 19.3 | 58.0 | 0.3 | 0.4 | 0.0 | 38535 |

Table 6.2.2.1. Faroe saithe (Division 5.b). Catch number-at-age by fleet categories in 2015 (calculated from gutted weights).

| Age | JIGGERS | Single <br> TRAWLERS $>1000 \mathrm{HP}$ | Pairtrawlers <1000 HP | Pairtrawlers $>1000 \mathrm{HP}$ | Others | Total Division 5.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 4 | 22 | 0 | 27 |
| 3 | 27 | 28 | 447 | 1233 | 45 | 1781 |
| 4 | 48 | 40 | 349 | 1414 | 41 | 1892 |
| 5 | 58 | 39 | 358 | 1645 | 49 | 2149 |
| 6 | 37 | 40 | 296 | 1206 | 30 | 1608 |
| 7 | 17 | 15 | 118 | 556 | 14 | 719 |
| 8 | 5 | 5 | 37 | 185 | 4 | 236 |
| 9 | 5 | 3 | 25 | 111 | 4 | 149 |
| 10 | 3 | 1 | 10 | 55 | 2 | 72 |
| 11 | 1 | 1 | 7 | 48 | 1 | 58 |
| 12 | 0 | 1 | 6 | 20 | 0 | 27 |
| 13 | 0 | 1 | 5 | 20 | 0 | 26 |
| 14 | 0 | 0 | 1 | 6 | 0 | 7 |
| 15 | 0 | 0 | 1 | 3 | 0 | 5 |
| Total No. | 202 | 174 | 1665 | 6525 | 192 | 8757 |
| Catch, t. | 498 | 428 | 3764 | 15818 | 449 | 20957 |

Table 6.2.2.2. Faroe saithe (Division 5.b). Catch number-at-age (thousands) from the commercial fleet (1961-2014)

| CN | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 183 | 379 | 483 | 403 | 216 | 129 | 116 | 82 | 45 | 27 | 6 | 49 |
| 1962 | 562 | 542 | 617 | 495 | 286 | 131 | 129 | 113 | 71 | 29 | 13 | 63 |
| 1963 | 614 | 340 | 340 | 415 | 406 | 202 | 174 | 158 | 94 | 169 | 61 | 44 |
| 1964 | 684 | 1908 | 1506 | 617 | 572 | 424 | 179 | 150 | 100 | 83 | 47 | 44 |
| 1965 | 996 | 850 | 1708 | 965 | 510 | 407 | 306 | 201 | 156 | 120 | 89 | 76 |
| 1966 | 488 | 1540 | 1201 | 1686 | 806 | 377 | 294 | 205 | 156 | 94 | 52 | 79 |
| 1967 | 595 | 796 | 1364 | 792 | 1192 | 473 | 217 | 190 | 97 | 75 | 38 | 27 |
| 1968 | 614 | 1689 | 1116 | 1095 | 548 | 655 | 254 | 128 | 89 | 59 | 40 | 88 |
| 1969 | 1191 | 2086 | 2294 | 1414 | 1118 | 589 | 580 | 239 | 115 | 100 | 36 | 54 |
| 1970 | 1445 | 6577 | 1558 | 1478 | 899 | 730 | 316 | 241 | 86 | 48 | 46 | 38 |
| 1971 | 2857 | 3316 | 5585 | 1005 | 828 | 469 | 326 | 164 | 100 | 54 | 13 | 33 |
| 1972 | 2714 | 1774 | 2588 | 2742 | 1529 | 1305 | 1017 | 743 | 330 | 133 | 28 | 49 |
| 1973 | 2515 | 6253 | 7075 | 3478 | 1634 | 693 | 550 | 403 | 215 | 103 | 25 | 58 |
| 1974 | 3504 | 4126 | 4011 | 2784 | 1401 | 640 | 368 | 340 | 197 | 124 | 45 | 96 |
| 1975 | 2062 | 3361 | 3801 | 1939 | 1045 | 714 | 302 | 192 | 193 | 126 | 64 | 108 |
| 1976 | 3178 | 3217 | 1720 | 1250 | 877 | 641 | 468 | 223 | 141 | 96 | 60 | 131 |
| 1977 | 1609 | 2937 | 2034 | 1288 | 767 | 708 | 498 | 338 | 272 | 129 | 80 | 121 |
| 1978 | 611 | 1743 | 1736 | 548 | 373 | 479 | 466 | 473 | 407 | 211 | 146 | 178 |
| 1979 | 287 | 933 | 1341 | 1033 | 584 | 414 | 247 | 473 | 368 | 206 | 136 | 349 |
| 1980 | 996 | 877 | 720 | 673 | 726 | 284 | 212 | 171 | 196 | 156 | 261 | 369 |
| 1981 | 411 | 1804 | 769 | 932 | 908 | 734 | 343 | 192 | 92 | 128 | 176 | 717 |
| 1982 | 387 | 4076 | 994 | 1114 | 380 | 417 | 296 | 105 | 88 | 56 | 49 | 797 |
| 1983 | 2483 | 1103 | 5052 | 1343 | 575 | 339 | 273 | 98 | 98 | 99 | 25 | 416 |
| 1984 | 368 | 11067 | 2359 | 4093 | 875 | 273 | 161 | 52 | 65 | 59 | 18 | 176 |
| 1985 | 1224 | 3990 | 5583 | 1182 | 1898 | 273 | 103 | 38 | 26 | 72 | 41 | 162 |
| 1986 | 1167 | 1997 | 4473 | 3730 | 953 | 1077 | 245 | 104 | 67 | 33 | 56 | 69 |
| 1987 | 1581 | 5793 | 3827 | 2785 | 990 | 532 | 333 | 81 | 43 | 5 | 11 | 81 |
| 1988 | 866 | 2950 | 9555 | 2784 | 1300 | 621 | 363 | 159 | 27 | 43 | 15 | 2 |
| 1989 | 451 | 5981 | 5300 | 7136 | 793 | 546 | 185 | 83 | 55 | 10 | 2 | 27 |
| 1990 | 294 | 3833 | 10120 | 9219 | 5070 | 477 | 123 | 61 | 60 | 18 | 19 | 42 |
| 1991 | 1030 | 5125 | 7452 | 5544 | 3487 | 1630 | 405 | 238 | 128 | 77 | 22 | 19 |
| 1992 | 521 | 4067 | 3667 | 2679 | 1373 | 894 | 613 | 123 | 63 | 37 | 52 | 19 |
| 1993 | 1316 | 2611 | 4689 | 1665 | 858 | 492 | 448 | 245 | 54 | 34 | 10 | 8 |
| 1994 | 690 | 3961 | 2663 | 2368 | 746 | 500 | 307 | 303 | 150 | 28 | 19 | 2 |
| 1995 | 398 | 1019 | 3468 | 1836 | 1177 | 345 | 241 | 192 | 104 | 73 | 25 | 19 |
| 1996 | 297 | 1087 | 1146 | 1449 | 1156 | 521 | 132 | 77 | 64 | 45 | 29 | 8 |
| 1997 | 344 | 832 | 2440 | 1767 | 1335 | 624 | 165 | 71 | 29 | 48 | 29 | 23 |
| 1998 | 163 | 1689 | 1934 | 3475 | 1379 | 683 | 368 | 77 | 32 | 28 | 24 | 21 |
| 1999 | 322 | 655 | 3096 | 2551 | 4113 | 915 | 380 | 147 | 24 | 27 | 5 | 37 |
| 2000 | 811 | 2830 | 1484 | 4369 | 2226 | 2725 | 348 | 186 | 56 | 18 | 2 | 5 |
| 2001 | 1125 | 2452 | 8437 | 2155 | 3680 | 1539 | 1334 | 293 | 90 | 24 | 19 | 13 |
| 2002 | 302 | 8399 | 5962 | 9786 | 862 | 1280 | 465 | 362 | 33 | 36 | 8 | 1 |


| $\mathbf{C N}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4 +} \mathbf{4}$ |
| ---: | :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2003 | 330 | 2432 | 11152 | 3994 | 4287 | 417 | 419 | 304 | 91 | 40 | 3 | 0 |
| 2004 | 76 | 2011 | 8544 | 8762 | 2125 | 1807 | 265 | 293 | 146 | 100 | 10 | 2 |
| 2005 | 454 | 2948 | 9486 | 16606 | 7099 | 843 | 810 | 32 | 102 | 27 | 3 | 0 |
| 2006 | 1475 | 5045 | 7781 | 7712 | 10296 | 3760 | 640 | 282 | 32 | 12 | 12 | 5 |
| 2007 | 831 | 3320 | 11305 | 6473 | 3781 | 4294 | 1538 | 406 | 81 | 11 | 9 | 3 |
| 2008 | 4784 | 3108 | 3598 | 9370 | 3594 | 2223 | 2048 | 444 | 159 | 12 | 6 | 0 |
| 2009 | 459 | 7412 | 4978 | 1842 | 5167 | 2009 | 1696 | 1069 | 292 | 41 | 3 | 1 |
| 2010 | 2324 | 2916 | 5298 | 1125 | 1009 | 2098 | 1248 | 832 | 376 | 51 | 22 | 0 |
| 2011 | 1897 | 2744 | 1940 | 1804 | 477 | 530 | 704 | 521 | 439 | 138 | 34 | 4 |
| 2012 | 859 | 9833 | 4142 | 1252 | 901 | 304 | 307 | 399 | 229 | 136 | 91 | 21 |
| 2013 | 721 | 5172 | 4219 | 2242 | 511 | 209 | 122 | 96 | 146 | 85 | 39 | 36 |
| 2014 | 879 | 2323 | 3143 | 1681 | 865 | 330 | 99 | 92 | 70 | 55 | 16 | 1 |
| 2015 | 2135 | 2269 | 2577 | 1928 | 863 | 283 | 179 | 86 | 69 | 33 | 31 | 15 |

Table 6.2.2.3. Faroe saithe (Division 5.b). Sampling intensity in 2001-2015.



Table 6.2.3.1. Faroe saithe (Division 5.b). Catch weights at age (kg)(equal to stock-weights) from the commercial fleet (1961-2015). The value for 2016 is used for short-term projections.

| CW | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 1.43 | 2.302 | 3.348 | 4.287 | 5.128 | 6.155 | 7.06 | 7.265 | 7.497 | 8.198 | 9.154 | 9.992 |
| 1962 | 1.273 | 2.045 | 3.293 | 4.191 | 5.146 | 5.655 | 6.469 | 6.706 | 7.15 | 7.903 | 8.449 | 9.658 |
| 1963 | 1.28 | 2.197 | 3.212 | 4.568 | 5.056 | 5.932 | 6.259 | 8 | 7.265 | 8.551 | 9.02 | 9.818 |
| 1964 | 1.175 | 2.055 | 3.266 | 4.255 | 5.038 | 5.694 | 6.662 | 6.837 | 7.686 | 8.348 | 8.123 | 9.423 |
| 1965 | 1.181 | 2.125 | 2.941 | 4.096 | 4.878 | 5.932 | 6.321 | 7.288 | 8.074 | 7.878 | 9.479 | 9.849 |
| 1966 | 1.361 | 2.026 | 3.055 | 3.658 | 4.585 | 5.52 | 6.837 | 7.265 | 7.662 | 8.123 | 10.21 | 9.883 |
| 1967 | 1.273 | 1.78 | 2.534 | 3.572 | 4.368 | 5.313 | 5.812 | 6.554 | 7.806 | 7.591 | 8.551 | 9.135 |
| 1968 | 1.302 | 1.737 | 2.036 | 3.12 | 4.049 | 5.183 | 6.238 | 7.52 | 8.049 | 8.654 | 8.298 | 9.748 |
| 1969 | 1.188 | 1.667 | 2.302 | 2.853 | 3.673 | 5.002 | 5.714 | 6.405 | 6.554 | 7.591 | 7.951 | 9.096 |
| 1970 | 1.244 | 1.445 | 2.249 | 2.853 | 3.515 | 4.418 | 5.444 | 5.733 | 6.662 | 7.31 | 9.047 | 9.634 |
| 1971 | 1.101 | 1.316 | 1.818 | 2.978 | 3.702 | 4.271 | 5.388 | 5.972 | 6.49 | 7.173 | 7.38 | 9.612 |
| 1972 | 1.043 | 1.485 | 2.055 | 2.829 | 3.791 | 4.175 | 4.808 | 5.294 | 6.948 | 6.727 | 7.591 | 9.609 |
| 1973 | 1.306 | 1.754 | 1.899 | 2.7 | 4.426 | 5.264 | 6.156 | 6.334 | 8.076 | 8.777 | 9.782 | 11.115 |
| 1974 | 1.615 | 1.723 | 2.493 | 2.824 | 3.524 | 5.197 | 6.279 | 6.454 | 7.07 | 7.773 | 8.763 | 10.83 |
| 1975 | 1.293 | 1.924 | 2.623 | 3.621 | 4.128 | 4.754 | 5.952 | 7.073 | 8.352 | 9.032 | 9.984 | 11.082 |
| 1976 | 1.162 | 1.79 | 3.074 | 3.291 | 4.579 | 4.648 | 5.116 | 6.314 | 7.069 | 7.069 | 7.808 | 9.714 |
| 1977 | 1.223 | 1.641 | 2.66 | 3.79 | 4.239 | 5.597 | 5.35 | 5.912 | 6.837 | 6.727 | 6.948 | 9.258 |
| 1978 | 1.493 | 2.324 | 3.068 | 3.746 | 4.913 | 4.368 | 5.276 | 5.832 | 6.053 | 6.706 | 7.686 | 8.516 |
| 1979 | 1.22 | 1.88 | 2.62 | 3.4 | 4.18 | 4.95 | 5.69 | 6.38 | 7.02 | 7.26 | 8.15 | 9.618 |
| 1980 | 1.23 | 2.12 | 3.32 | 4.28 | 5.16 | 6.42 | 6.87 | 7.09 | 7.93 | 8.07 | 8.59 | 10.142 |
| 1981 | 1.31 | 2.13 | 3 | 3.81 | 4.75 | 5.25 | 5.95 | 6.43 | 7 | 7.47 | 8.14 | 9.43 |
| 1982 | 1.337 | 1.851 | 2.951 | 3.577 | 4.927 | 6.243 | 7.232 | 7.239 | 8.346 | 8.345 | 8.956 | 10.227 |
| 1983 | 1.208 | 2.029 | 2.965 | 4.143 | 4.724 | 5.901 | 6.811 | 7.051 | 7.248 | 8.292 | 9.478 | 10.509 |
| 1984 | 1.431 | 1.953 | 2.47 | 3.85 | 5.177 | 6.347 | 7.825 | 6.746 | 8.636 | 8.467 | 8.556 | 10.802 |
| 1985 | 1.401 | 2.032 | 2.965 | 3.596 | 5.336 | 7.202 | 6.966 | 9.862 | 10.67 | 10.46 | 10.202 | 13.055 |
| 1986 | 1.718 | 1.986 | 2.618 | 3.277 | 4.186 | 5.589 | 6.05 | 6.15 | 9.536 | 9.823 | 7.303 | 12.773 |
| 1987 | 1.609 | 1.835 | 2.395 | 3.182 | 4.067 | 5.149 | 5.501 | 6.626 | 6.343 | 10.245 | 8.491 | 10.482 |
| 1988 | 1.5 | 1.975 | 1.978 | 2.937 | 3.798 | 4.419 | 5.115 | 6.712 | 9.04 | 9.364 | 9.142 | 10.216 |
| 1989 | 1.309 | 1.735 | 1.907 | 2.373 | 3.81 | 4.667 | 5.509 | 5.972 | 6.939 | 8.543 | 9.514 | 10.484 |
| 1990 | 1.223 | 1.633 | 1.83 | 2.052 | 2.866 | 4.474 | 5.424 | 6.469 | 6.343 | 8.418 | 7.383 | 8.64 |
| 1991 | 1.24 | 1.568 | 1.864 | 2.211 | 2.648 | 3.38 | 4.816 | 5.516 | 6.407 | 7.395 | 8.079 | 8.674 |
| 1992 | 1.264 | 1.602 | 2.069 | 2.554 | 3.057 | 4.078 | 5.012 | 6.768 | 7.754 | 8.303 | 7.786 | 9.301 |
| 1993 | 1.408 | 1.86 | 2.323 | 3.131 | 3.73 | 4.394 | 5.209 | 6.54 | 8.403 | 7.275 | 9.414 | 9.64 |


| CW | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | 10 | 11 | 12 | 13 | $14+$ |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| 1994 | 1.503 | 1.951 | 2.267 | 2.936 | 4.214 | 4.971 | 5.657 | 5.95 | 6.891 | 8.752 | 9.752 | 7.989 |
| 1995 | 1.456 | 2.177 | 2.42 | 2.895 | 3.651 | 5.064 | 5.44 | 6.167 | 7.08 | 7.736 | 7.295 | 7.104 |
| 1996 | 1.432 | 1.875 | 2.496 | 3.229 | 3.744 | 4.964 | 6.375 | 6.745 | 7.466 | 7.284 | 8.47 | 10.125 |
| 1997 | 1.476 | 1.783 | 2.032 | 2.778 | 3.598 | 4.766 | 5.982 | 7.658 | 7.882 | 8.539 | 9.488 | 10.413 |
| 1998 | 1.388 | 1.711 | 1.954 | 2.405 | 3.3 | 4.22 | 4.999 | 6.391 | 6.665 | 8.214 | 8.485 | 8.845 |
| 1999 | 1.374 | 1.712 | 1.905 | 2.396 | 2.845 | 4.124 | 5.256 | 5.526 | 6.956 | 8.03 | 8.349 | 8.907 |
| 2000 | 1.477 | 1.606 | 2.077 | 2.36 | 2.977 | 3.48 | 4.851 | 5.268 | 6.523 | 4.727 | 8.807 | 8.972 |
| 2001 | 1.33 | 1.59 | 1.785 | 2.586 | 3.059 | 3.871 | 4.374 | 5.565 | 6.703 | 5.776 | 7.745 | 7.773 |
| 2002 | 1.142 | 1.46 | 1.652 | 1.969 | 3.13 | 3.589 | 4.513 | 5.138 | 6.422 | 8.026 | 4.759 | 11.357 |
| 2003 | 1.123 | 1.304 | 1.614 | 1.977 | 2.532 | 3.97 | 4.834 | 5.499 | 6.099 | 6.987 | 5.961 | 10 |
| 2004 | 1.143 | 1.333 | 1.45 | 1.789 | 2.56 | 3.159 | 4.154 | 5.167 | 6.015 | 6.186 | 7.056 | 9.391 |
| 2005 | 1.148 | 1.325 | 1.516 | 1.672 | 2.087 | 2.975 | 3.79 | 6.087 | 6.134 | 6.651 | 7.424 | 10 |
| 2006 | 1.126 | 1.218 | 1.462 | 1.79 | 2.035 | 2.436 | 3.861 | 4.222 | 5.149 | 6.437 | 6.905 | 5.365 |
| 2007 | 1.058 | 1.391 | 1.413 | 1.824 | 2.361 | 2.682 | 3.278 | 4.104 | 4.998 | 6.331 | 7.844 | 7.971 |
| 2008 | 1.146 | 1.312 | 1.672 | 1.816 | 2.395 | 2.902 | 3.1 | 3.728 | 4.769 | 6.072 | 6.451 | 10 |
| 2009 | 0.938 | 1.485 | 1.893 | 2.411 | 2.601 | 3.147 | 3.634 | 4.024 | 5.014 | 5.828 | 6.308 | 9.011 |
| 2010 | 1.429 | 1.706 | 2.166 | 2.551 | 3.172 | 3.411 | 3.972 | 4.352 | 5.083 | 4.941 | 5.305 | 10 |
| 2011 | 1.111 | 1.693 | 2.253 | 2.918 | 3.609 | 4.204 | 4.531 | 5.087 | 5.416 | 6.087 | 6.763 | 7.916 |
| 2012 | 1.029 | 1.334 | 1.626 | 2.709 | 3.785 | 4.448 | 4.799 | 5.207 | 5.562 | 6.018 | 7.143 | 6.247 |
| 2013 | 1.208 | 1.466 | 1.778 | 2.069 | 3.553 | 4.292 | 5.191 | 5.742 | 5.919 | 6.417 | 7.941 | 7.138 |
| 2014 | 1.369 | 1.724 | 2.163 | 2.868 | 3.325 | 5.903 | 5.899 | 6.877 | 6.784 | 7.467 | 7.121 | 11.31 |
| 2015 | 0.932 | 1.555 | 2.091 | 3.17 | 4.208 | 5.032 | 6.715 | 7.858 | 7.428 | 7.565 | 7.629 | 9.367 |
| 2016 | 1.295 | 1.120 | 1.997 | 2.719 | 4.076 | 5.373 | 5.935 | 6.826 | 6.710 | 7.150 | 7.564 | 9.272 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 6.2.4.1. Faroe saithe (Division 5.b). Proportion mature at age (1982-2015). Maturities-at-age from 1961 to 1981 are fixed and equal to those in 1982.

| MAT | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0.03 | 0.22 | 0.52 | 0.79 | 0.92 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1983 | 0.03 | 0.27 | 0.61 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1984 | 0.04 | 0.28 | 0.60 | 0.88 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1985 | 0.05 | 0.29 | 0.59 | 0.85 | 0.97 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1986 | 0.05 | 0.28 | 0.57 | 0.82 | 0.94 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1987 | 0.05 | 0.27 | 0.55 | 0.79 | 0.92 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1988 | 0.05 | 0.26 | 0.53 | 0.77 | 0.90 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1989 | 0.04 | 0.23 | 0.51 | 0.76 | 0.89 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1990 | 0.03 | 0.19 | 0.49 | 0.75 | 0.89 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1991 | 0.03 | 0.17 | 0.48 | 0.75 | 0.88 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1992 | 0.02 | 0.17 | 0.48 | 0.75 | 0.89 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1993 | 0.02 | 0.17 | 0.49 | 0.77 | 0.91 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1994 | 0.01 | 0.17 | 0.49 | 0.78 | 0.93 | 1.00 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1995 | 0.01 | 0.17 | 0.49 | 0.78 | 0.93 | 1.00 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1996 | 0.01 | 0.17 | 0.47 | 0.75 | 0.90 | 1.00 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1997 | 0.01 | 0.16 | 0.44 | 0.70 | 0.87 | 0.98 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1998 | 0.02 | 0.16 | 0.41 | 0.64 | 0.83 | 0.96 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1999 | 0.02 | 0.16 | 0.38 | 0.60 | 0.79 | 0.94 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2000 | 0.02 | 0.16 | 0.37 | 0.58 | 0.77 | 0.92 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2001 | 0.01 | 0.17 | 0.37 | 0.56 | 0.75 | 0.91 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2002 | 0.01 | 0.17 | 0.37 | 0.56 | 0.74 | 0.89 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2003 | 0.01 | 0.18 | 0.37 | 0.56 | 0.74 | 0.88 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2004 | 0.01 | 0.18 | 0.38 | 0.57 | 0.74 | 0.88 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2005 | 0.00 | 0.18 | 0.39 | 0.59 | 0.76 | 0.89 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2006 | 0.00 | 0.18 | 0.40 | 0.62 | 0.78 | 0.90 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2007 | 0.00 | 0.19 | 0.42 | 0.64 | 0.80 | 0.91 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2008 | 0.01 | 0.20 | 0.43 | 0.66 | 0.82 | 0.92 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2009 | 0.01 | 0.21 | 0.45 | 0.68 | 0.84 | 0.94 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2010 | 0.02 | 0.23 | 0.47 | 0.71 | 0.87 | 0.95 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2011 | 0.03 | 0.24 | 0.49 | 0.72 | 0.88 | 0.96 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2012 | 0.03 | 0.25 | 0.50 | 0.73 | 0.89 | 0.97 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2013 | 0.04 | 0.25 | 0.50 | 0.74 | 0.90 | 0.97 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2014 | 0.04 | 0.26 | 0.51 | 0.74 | 0.90 | 0.98 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2015 | 0.05 | 0.26 | 0.51 | 0.74 | 0.9 | 0.98 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

Table 6.3.1. Faroe saithe (Division 5.b). Effort (hours) and catch in number-at-age for the commercial pairtrawlers (1995-2015)

| YEAR | EFFORT | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 10883 | 47 | 180 | 577 | 236 | 146 | 49 | 24 | 19 | 14 |
| 1996 | 47531 | 310 | 958 | 821 | 1119 | 503 | 282 | 133 | 127 | 70 |
| 1997 | 34606 | 199 | 533 | 1488 | 1013 | 768 | 333 | 73 | 33 | 10 |
| 1998 | 34144 | 107 | 656 | 1148 | 1486 | 730 | 325 | 170 | 40 | 13 |
| 1999 | 43218 | 174 | 487 | 1554 | 2016 | 2024 | 817 | 190 | 83 | 12 |
| 2000 | 43920 | 434 | 1566 | 913 | 2700 | 1333 | 1604 | 192 | 106 | 31 |
| 2001 | 41534 | 611 | 1438 | 4946 | 1165 | 1855 | 748 | 618 | 127 | 29 |
| 2002 | 41575 | 133 | 3976 | 3964 | 6888 | 520 | 682 | 246 | 177 | 25 |
| 2003 | 38076 | 141 | 1494 | 6560 | 2373 | 2263 | 197 | 212 | 124 | 35 |
| 2004 | 35237 | 43 | 1200 | 5089 | 5116 | 1035 | 762 | 113 | 116 | 53 |
| 2005 | 32493 | 188 | 1189 | 4039 | 7266 | 3130 | 320 | 291 | 7 | 43 |
| 2006 | 25068 | 140 | 1176 | 2410 | 2584 | 3700 | 1376 | 268 | 85 | 14 |
| 2007 | 24885 | 204 | 879 | 2913 | 1815 | 1034 | 1215 | 435 | 110 | 19 |
| 2008 | 25014 | 796 | 762 | 947 | 2641 | 1063 | 726 | 611 | 156 | 51 |
| 2009 | 67648 | 154 | 4082 | 3377 | 1283 | 3612 | 1402 | 1153 | 751 | 195 |
| 2010 | 61407 | 459 | 2019 | 3586 | 737 | 657 | 1325 | 814 | 518 | 245 |
| 2011 | 58209 | 397 | 1936 | 1367 | 1257 | 323 | 356 | 488 | 366 | 310 |
| 2012 | 58244 | 366 | 5652 | 2332 | 756 | 554 | 187 | 189 | 252 | 143 |
| 2013 | 43770 | 424 | 3047 | 2462 | 1295 | 293 | 122 | 71 | 56 | 83 |
| 2014 | 48449 | 625 | 1624 | 2226 | 1200 | 613 | 216 | 72 | 70 | 50 |
| 2015 | 37639 | 437 | 1414 | 1645 | 1206 | 556 | 185 | 111 | 55 | 48 |

Table 6.3.2. Faroe saithe (Division 5.b). Diagnostics from XSA with commercial pairtrawler tuning series (spaly)

FLR XSA Diagnostics 2016-04-12 10:28:36
cpue data from indices
Catch data for 55 years 1961 to 2015. Ages 3 to 14 .
fleet first age last age first year last year alpha beta 1 PairTrawlers_GLM_SD 31119952015 <NA> <NA>

Time-series weights:
Tapered time weighting not applied
Catchability analysis :

Catchability independent of size for all ages
Catchability independent of age for ages > 8
Terminal population estimation :
Survivor estimates shrunk towards the mean F
of the final 5 years or the 3 oldest ages.
S.E. of the mean to which the estimates are shrunk $=2$

Minimum standard error for population
estimates derived from each fleet $=0.3$
prior weighting not applied
Regression weights
year
age 2006200720082009201020112012201320142015
$\begin{array}{lllllllllll}\text { all } & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$
Fishing mortalities
year
age 2006200720082009201020112012201320142015
30.0760 .0490 .1820 .0370 .1190 .0660 .0350 .0300 .0240 .038
40.1030 .2450 .2580 .4760 .3430 .2020 .5600 .3010 .1290 .080
50.2960 .3530 .4580 .8600 .7600 .4040 .5320 .5000 .3020 .206
60.4140 .4320 .5580 .4510 .4720 .6410 .4980 .6240 .3800 .307

```
70.6190.3670.456 0.701 0.480 0.374 0.795 0.388 0.5240.342
8 0.715 0.574 0.3840.501 0.700 0.503 0.436 0.422 0.469 0.322
90.4880.7370.600 0.5720.6810.5370.620 0.312 0.3620.505
100.9080.6670.4850.7430.6210.6880.678 0.398 0.4110.621
1 1 0 . 5 1 8 0 . 7 3 1 0 . 6 0 5 0 . 6 9 6 ~ 0 . 6 4 1 0 . 8 0 8 ~ 0 . 7 5 8 ~ 0 . 5 6 9 ~ 0 . 5 7 1 0 . 6 2 6 ~
1 2 0 . 1 2 0 0 . 3 3 5 0 . 2 1 7 0 . 3 0 3 0 . 2 4 1 0 . 5 1 5 0 . 6 3 7 0 . 7 2 1 ~ 0 . 4 3 4 0 . 5 8 6 ~
130.2240.124 0.308 0.077 0.264 0.251 0.783 0.374 0.278 0.469
1 4 0 . 2 2 4 0 . 1 2 4 0 . 3 0 8 ~ 0 . 0 7 7 ~ 0 . 2 6 4 ~ 0 . 2 5 1 ~ 0 . 7 8 3 ~ 0 . 3 7 4 ~ 0 . 2 7 8 ~ 0 . 4 6 9 )
XSA population number (Thousand)
    age
year 
20062226456937 335242513224659813618335228711766 27
2007193451689442051204071359810873 3259 922173 43 85 28
200831701150861082824200 108517712501712773876825 0
20091406821626 9539560911335563243032254644173 45 15
201022830111021099933062926460527931988 878 263105 0
20113304516589645142111689148218721158875 379169 20
2012277882533811099 3526 1816 951 734 89547631918542
2 0 1 3 2 6 7 9 9 2 1 9 7 4 1 1 8 4 8 ~ 5 3 3 9 1 7 5 4 ~ 6 7 1 ~ 5 0 3 ~ 3 2 3 ~ 3 7 2 1 8 3 1 3 8 1 2 7 )
201440622 21289133115883 2343 974 360 302 178 173 73 5
20156283632463153288054 32951135499206164829144
Estimated population abundance at 1st Jan 2016
```

    age
    $\begin{array}{lllllllllll}\text { year } & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 1011 & 12 & 1314\end{array}$
2016474951524526102184850191767324690723747

Fleet: PairTrawlers_GLM_SD

Log catchability residuals.

```
    year
age 19095}191996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007 2008 2009
2010 2011 2012 2013 2014 2015
3 -0.441 0.440-0.006 0.355-0.926 0.475 -0.031-1.746-1.102-2.037-0.737 0.407 0.918 1.844-0.050
0.694 0.207 0.284 0.750 0.618 0.083
4 -0.057 -0.755 -0.550-0.638-0.202 -0.588-0.090 0.036 -1.101 -0.738-0.478 -0.470 0.528 0.499 0.921
0.921 0.466 1.275 0.969 0.190-0.141
5 0.437-0.670-0.690-0.441 -0.656 -0.208 0.023 0.381 0.062 -0.479 -0.050-0.100 -0.104 0.171 0.747
0.720 0.187 0.235 0.496 0.088-0.148
6 -0.196 -0.175-0.078-0.665-0.053 0.009 0.344 0.650}0.206 0.050 0.096 -0.049 -0.179 0.077-0.225-
0.145 0.274-0.120 0.344-0.038-0.128
7 0.164-0.398 0.231 0.066 -0.164-0.031 0.331 0.214 0.375 -0.006 0.179 0.295-0.489-0.201 0.091-
0.259-0.413 0.237-0.257 0.151-0.117
8 0.138 0.198 0.145 0.029 0.608 0.315 0.153 0.182 0.043 0.208-0.545 0.391 -0.076-0.337-0.307 0.021
-0.192 -0.422 -0.222 -0.103-0.225
9 0.005 0.435 0.041 0.286 0.021-0.078 0.453-0.158-0.123 0.523 0.315 0.147 0.171 0.016-0.202 0.025
-0.095-0.072 -0.525-0.256 0.169
    10-0.344 1.103 0.101 0.231 0.218 0.287 0.564 0.336-0.002 0.124-1.292 0.433 0.030-0.031 0.089-
0.113 0.163 0.042-0.281-0.085 0.404
\(11-0.0410 .147-0.363-0.043-0.526 \quad 0.052 \quad 0.081-0.007-0.293 \quad 0.137 \quad 0.068 \quad 0.251-0.0240 .096-0.026-\) 0.0360 .3280 .1400 .0480 .1790 .496
```

Mean $\log$ catchability and standard error of ages with catchability
independent of year-class strength and constant w.r.t. time

| 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Mean_Logq -15.4706-13.3917-12.4239-12.0639-11.9395-11.8752-11.8752-11.8752-11.8752
$\begin{array}{llllllllll}\text { S.E_Logq } & 0.4616 & 0.4616 & 0.4616 & 0.4616 & 0.4616 & 0.4616 & 0.4616 & 0.4616 & 0.4616\end{array}$

Terminal year survivor and F summaries:
,Age 3 Year class =2012
source
scaledWts survivors yrcls

PairTrawlers_GLM_SD 0.816538092012
fshk $\quad 0.184 \quad 342012012$
,Age 4 Year class =2011
source

| scaledWts survivors yrcls |  |  |
| :---: | :---: | :---: |
| PairTrawlers_GLM_SD | 0.887 | 213032011 |
| fshk 0.113 | 5676201 |  |
| ,Age 5 Year class =2010 |  |  |
| source |  |  |
| scaledWts survivors yrcls |  |  |
| PairTrawlers_GLM_SD | 0.944 | 88132010 |
| fshk 0.056 | 3562201 |  |
| ,Age 6 Year class =2009 |  |  |
| source |  |  |
| scaledWts survivors yrcls |  |  |
| PairTrawlers_GLM_SD | 0.97 | 42682009 |
| fshk 0.03 2 | 2513200 |  |
| ,Age 7 Year class =2008 |  |  |
| source |  |  |
| scaledWts survivors yrcls |  |  |
| PairTrawlers_GLM_SD | 0.969 | 17052008 |
| fshk 0.031 | 1155200 |  |
| ,Age 8 Year class =2007 |  |  |
| source |  |  |
| scaledWts survivors yrcls |  |  |
| PairTrawlers_GLM_SD | 0.97 | 5382007 |
| fshk 0.03 | 3852007 |  |
| ,Age 9 Year class =2006 |  |  |
| source |  |  |
| scaledWts survivors yrcls |  |  |
| PairTrawlers_GLM_SD | 0.964 | 2922006 |
| fshk 0.036 | 246200 |  |

,Age 10 Year class =2005
source

| PairTrawlers_GLM_SD | 0.909 | 1352005 |
| :---: | :---: | :---: |
| fshk 0.091 | 103200 |  |
| ,Age 11 Year class =2004 |  |  |
| source |  |  |
| scaledWts survivors yrcls |  |  |
| PairTrawlers_GLM_SD | 0.96 | 1182004 |
| fshk 0.04 | 652004 |  |
| ,Age 12 Year class =2003 |  |  |
| source |  |  |
| scaledWts survivors yrcls |  |  |
| fshk 1 44 2003 |  |  |
| ,Age 13 Year class =2002 |  |  |
| source |  |  |
| scaledWts survivors yrcls |  |  |

fshk $1 \quad 332002$

Table 6.5.1. Faroe saithe (Division 5.b). Fishing mortality-at-age (1961-2015). The value for 2016 is used for short-term prognosis.

| F | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 0.026 | 0.058 | 0.109 | 0.143 | 0.12 | 0.1 | 0.11 | 0.106 | 0.112 | 0.181 | 0.134 | 0.134 |
| 1962 | 0.052 | 0.101 | 0.127 | 0.156 | 0.143 | 0.099 | 0.138 | 0.149 | 0.125 | 0.098 | 0.124 | 0.124 |
| 1963 | 0.035 | 0.04 | 0.085 | 0.118 | 0.185 | 0.142 | 0.185 | 0.25 | 0.178 | 0.491 | 0.308 | 0.308 |
| 1964 | 0.052 | 0.144 | 0.251 | 0.218 | 0.236 | 0.301 | 0.18 | 0.241 | 0.248 | 0.235 | 0.243 | 0.243 |
| 1965 | 0.05 | 0.085 | 0.186 | 0.253 | 0.283 | 0.263 | 0.37 | 0.316 | 0.424 | 0.532 | 0.427 | 0.427 |
| 1966 | 0.026 | 0.103 | 0.167 | 0.283 | 0.348 | 0.35 | 0.308 | 0.456 | 0.433 | 0.493 | 0.464 | 0.464 |
| 1967 | 0.027 | 0.053 | 0.125 | 0.158 | 0.332 | 0.354 | 0.349 | 0.335 | 0.407 | 0.384 | 0.378 | 0.378 |
| 1968 | 0.03 | 0.099 | 0.098 | 0.14 | 0.156 | 0.307 | 0.326 | 0.358 | 0.258 | 0.467 | 0.363 | 0.363 |
| 1969 | 0.034 | 0.136 | 0.189 | 0.175 | 0.207 | 0.25 | 0.493 | 0.586 | 0.639 | 0.518 | 0.586 | 0.586 |
| 1970 | 0.044 | 0.262 | 0.142 | 0.179 | 0.16 | 0.202 | 0.206 | 0.39 | 0.431 | 0.609 | 0.48 | 0.48 |
| 1971 | 0.086 | 0.135 | 0.373 | 0.128 | 0.144 | 0.117 | 0.13 | 0.157 | 0.277 | 0.534 | 0.325 | 0.325 |
| 1972 | 0.094 | 0.07 | 0.148 | 0.316 | 0.293 | 0.354 | 0.4 | 0.49 | 0.541 | 0.73 | 0.592 | 0.592 |
| 1973 | 0.125 | 0.325 | 0.438 | 0.304 | 0.315 | 0.209 | 0.246 | 0.272 | 0.253 | 0.32 | 0.283 | 0.283 |
| 1974 | 0.222 | 0.311 | 0.358 | 0.307 | 0.192 | 0.195 | 0.164 | 0.237 | 0.207 | 0.227 | 0.225 | 0.225 |
| 1975 | 0.141 | 0.345 | 0.528 | 0.293 | 0.18 | 0.141 | 0.132 | 0.12 | 0.205 | 0.198 | 0.175 | 0.175 |
| 1976 | 0.196 | 0.34 | 0.298 | 0.328 | 0.208 | 0.16 | 0.129 | 0.137 | 0.122 | 0.149 | 0.136 | 0.136 |
| 1977 | 0.146 | 0.281 | 0.376 | 0.382 | 0.344 | 0.259 | 0.179 | 0.13 | 0.246 | 0.156 | 0.178 | 0.178 |
| 1978 | 0.085 | 0.233 | 0.267 | 0.163 | 0.18 | 0.375 | 0.272 | 0.259 | 0.228 | 0.307 | 0.266 | 0.266 |
| 1979 | 0.037 | 0.18 | 0.283 | 0.251 | 0.261 | 0.31 | 0.338 | 0.49 | 0.329 | 0.172 | 0.333 | 0.333 |
| 1980 | 0.088 | 0.153 | 0.205 | 0.224 | 0.281 | 0.195 | 0.258 | 0.415 | 0.386 | 0.226 | 0.344 | 0.344 |
| 1981 | 0.014 | 0.227 | 0.194 | 0.447 | 0.533 | 0.512 | 0.383 | 0.394 | 0.412 | 0.471 | 0.429 | 0.429 |
| 1982 | 0.028 | 0.184 | 0.188 | 0.477 | 0.329 | 0.502 | 0.399 | 0.191 | 0.315 | 0.477 | 0.33 | 0.33 |
| 1983 | 0.07 | 0.103 | 0.366 | 0.419 | 0.486 | 0.552 | 0.736 | 0.221 | 0.275 | 0.711 | 0.405 | 0.405 |
| 1984 | 0.016 | 0.498 | 0.332 | 0.575 | 0.535 | 0.451 | 0.558 | 0.292 | 0.224 | 0.265 | 0.262 | 0.262 |
| 1985 | 0.062 | 0.235 | 0.507 | 0.276 | 0.579 | 0.314 | 0.304 | 0.243 | 0.232 | 0.415 | 0.298 | 0.298 |
| 1986 | 0.021 | 0.137 | 0.452 | 0.774 | 0.375 | 0.785 | 0.518 | 0.578 | 0.895 | 0.518 | 0.67 | 0.67 |
| 1987 | 0.037 | 0.138 | 0.423 | 0.57 | 0.476 | 0.372 | 0.598 | 0.32 | 0.503 | 0.141 | 0.323 | 0.323 |
| 1988 | 0.022 | 0.089 | 0.355 | 0.631 | 0.576 | 0.629 | 0.47 | 0.649 | 0.167 | 1.598 | 0.813 | 0.813 |
| 1989 | 0.018 | 0.203 | 0.228 | 0.492 | 0.365 | 0.51 | 0.383 | 0.183 | 0.488 | 0.086 | 0.254 | 0.254 |
| 1990 | 0.016 | 0.203 | 0.626 | 0.783 | 0.8 | 0.391 | 0.202 | 0.208 | 0.196 | 0.29 | 0.232 | 0.232 |
| 1991 | 0.047 | 0.414 | 0.767 | 0.872 | 0.797 | 0.657 | 0.686 | 0.754 | 0.901 | 0.414 | 0.696 | 0.696 |
| 1992 | 0.03 | 0.262 | 0.595 | 0.707 | 0.547 | 0.48 | 0.556 | 0.455 | 0.452 | 0.726 | 0.549 | 0.549 |
| 1993 | 0.063 | 0.205 | 0.547 | 0.6 | 0.514 | 0.383 | 0.474 | 0.452 | 0.37 | 0.473 | 0.435 | 0.435 |
| 1994 | 0.046 | 0.274 | 0.333 | 0.596 | 0.598 | 0.651 | 0.44 | 0.695 | 0.557 | 0.333 | 0.533 | 0.533 |
| 1995 | 0.011 | 0.089 | 0.41 | 0.404 | 0.683 | 0.62 | 0.776 | 0.549 | 0.546 | 0.585 | 0.565 | 0.565 |
| 1996 | 0.014 | 0.039 | 0.137 | 0.3 | 0.484 | 0.755 | 0.513 | 0.612 | 0.353 | 0.484 | 0.487 | 0.487 |
| 1997 | 0.011 | 0.048 | 0.115 | 0.324 | 0.5 | 0.528 | 0.574 | 0.58 | 0.491 | 0.491 | 0.674 | 0.674 |
| 1998 | 0.014 | 0.071 | 0.15 | 0.238 | 0.454 | 0.52 | 0.694 | 0.583 | 0.567 | 1.376 | 0.49 | 0.49 |
| 1999 | 0.006 | 0.073 | 0.18 | 0.301 | 0.492 | 0.626 | 0.622 | 0.671 | 0.358 | 1.541 | 1.037 | 1.037 |
| 2000 | 0.025 | 0.068 | 0.234 | 0.417 | 0.469 | 0.721 | 0.518 | 0.724 | 0.588 | 0.502 | 0.403 | 0.403 |
| 2001 | 0.014 | 0.099 | 0.294 | 0.632 | 0.759 | 0.703 | 0.999 | 1.197 | 0.99 | 0.543 | 1.838 | 1.838 |
| 2002 | 0.003 | 0.14 | 0.371 | 0.66 | 0.563 | 0.659 | 0.472 | 0.841 | 0.382 | 1.753 | 0.348 | 0.348 |
| 2003 | 0.006 | 0.032 | 0.279 | 0.458 | 0.694 | 0.593 | 0.467 | 0.656 | 0.519 | 1.17 | 0.665 | 0.665 |
| 2004 | 0.002 | 0.043 | 0.148 | 0.369 | 0.474 | 0.726 | 0.988 | 0.711 | 0.786 | 2.449 | 1.134 | 1.134 |
| 2005 | 0.007 | 0.077 | 0.294 | 0.476 | 0.582 | 0.348 | 0.877 | 0.286 | 0.581 | 0.315 | 0.49 | 0.49 |


| F | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | 9 | 10 | $\mathbf{1 1}$ | 12 | 13 | $\mathbf{1 4 +}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 0.076 | 0.103 | 0.296 | 0.414 | 0.619 | 0.715 | 0.488 | 0.908 | 0.518 | 0.12 | 0.224 | 0.224 |
| 2007 | 0.049 | 0.245 | 0.353 | 0.432 | 0.367 | 0.574 | 0.737 | 0.667 | 0.731 | 0.335 | 0.124 | 0.124 |
| 2008 | 0.182 | 0.258 | 0.458 | 0.558 | 0.456 | 0.384 | 0.6 | 0.485 | 0.605 | 0.217 | 0.308 | 0.308 |
| 2009 | 0.037 | 0.476 | 0.86 | 0.451 | 0.701 | 0.501 | 0.572 | 0.743 | 0.696 | 0.303 | 0.077 | 0.077 |
| 2010 | 0.119 | 0.343 | 0.76 | 0.472 | 0.48 | 0.7 | 0.681 | 0.621 | 0.641 | 0.241 | 0.264 | 0.264 |
| 2011 | 0.066 | 0.202 | 0.404 | 0.641 | 0.374 | 0.503 | 0.537 | 0.688 | 0.808 | 0.515 | 0.251 | 0.251 |
| 2012 | 0.035 | 0.56 | 0.532 | 0.498 | 0.795 | 0.436 | 0.62 | 0.678 | 0.758 | 0.637 | 0.783 | 0.783 |
| 2013 | 0.03 | 0.301 | 0.5 | 0.624 | 0.388 | 0.422 | 0.312 | 0.398 | 0.569 | 0.721 | 0.374 | 0.374 |
| 2014 | 0.024 | 0.129 | 0.302 | 0.38 | 0.524 | 0.469 | 0.362 | 0.411 | 0.571 | 0.434 | 0.278 | 0.278 |
| 2015 | 0.038 | 0.08 | 0.206 | 0.307 | 0.342 | 0.322 | 0.505 | 0.621 | 0.626 | 0.586 | 0.469 | 0.469 |
| 2016 | 0.022 | 0.121 | 0.239 | 0.311 | 0.298 | 0.288 | 0.280 | 0.340 | 0.419 | 1.00 | 1.00 | 1.00 |

Table 6.3.2. Faroe saithe (Division 5.b). Stock number-at-age (start of year) (Thousands)(1961-2015). The value for 2016 is used for short-term prognosis.

| YEAR | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 7827.26 | 7421.86 | 5158.38 | 3351.65 | 2113.91 | 1494.26 | 1232.82 | 904.51 | 468.22 | 179.78 | 53.02 | 431.33 |
| 1962 | 12256.26 | 6242.83 | 5733.57 | 3786.29 | 2379.45 | 1535.28 | 1106.68 | 904.39 | 666.35 | 342.63 | 122.76 | 592.7 |
| 1963 | 19837.08 | 9526.05 | 4620.77 | 4135.96 | 2652.05 | 1689.34 | 1138.44 | 789.35 | 638.21 | 481.32 | 254.28 | 182.18 |
| 1964 | 14811.8 | 15685.65 | 7491.63 | 3475.53 | 3010.73 | 1803.95 | 1200.34 | 774.64 | 503.3 | 437.46 | 241.15 | 224.48 |
| 1965 | 22362.95 | 11507.97 | 11115.9 | 4770.94 | 2287.23 | 1947.41 | 1093.3 | 820.79 | 498.49 | 321.58 | 283.06 | 239.61 |
| 1966 | 21229.3 | 17408.01 | 8652.81 | 7555.46 | 3032.95 | 1411.16 | 1226.14 | 618.24 | 490.13 | 266.98 | 154.71 | 232.85 |
| 1967 | 24897.69 | 16939.52 | 12859.03 | 5997.62 | 4660.34 | 1753.87 | 814.24 | 737.85 | 320.68 | 260.13 | 133.53 | 94.13 |
| 1968 | 22879.44 | 19846.12 | 13148.65 | 9293.88 | 4193.8 | 2736.99 | 1007.96 | 470.29 | 432.19 | 174.78 | 145.12 | 316.81 |
| 1969 | 39798.62 | 18176.53 | 14720.36 | 9755.41 | 6618.39 | 2937.74 | 1648.19 | 595.42 | 269.22 | 273.31 | 89.71 | 133.05 |
| 1970 | 37092.28 | 31506.69 | 12994.19 | 9976.32 | 6707.61 | 4407.07 | 1872.27 | 824.62 | 271.23 | 116.37 | 133.29 | 109.05 |
| 1971 | 38446.77 | 29061.1 | 19844.38 | 9229.01 | 6830.57 | 4678.28 | 2947.67 | 1246.96 | 457.08 | 144.25 | 51.84 | 130.67 |
| 1972 | 33424.52 | 28892.43 | 20792.77 | 11193.69 | 6646.71 | 4843.19 | 3405.88 | 2118.37 | 872.53 | 283.74 | 69.24 | 119.79 |
| 1973 | 23621.9 | 24909.95 | 22049.94 | 14681.96 | 6683.55 | 4058.37 | 2784.46 | 1868.28 | 1062.08 | 415.77 | 111.96 | 258.1 |
| 1974 | 19420.68 | 17064.31 | 14736.6 | 11651.24 | 8873.55 | 3993.53 | 2695.66 | 1782.06 | 1164.97 | 675.02 | 247.21 | 524.53 |
| 1975 | 17327.33 | 12729.76 | 10237.71 | 8436 | 7020.16 | 5997.37 | 2690.53 | 1874.04 | 1151.38 | 775.54 | 440.46 | 739.88 |
| 1976 | 19709.34 | 12320.65 | 7381.08 | 4942.64 | 5152.33 | 4802.07 | 4264.18 | 1929.56 | 1360.61 | 768.04 | 520.95 | 1132.95 |
| 1977 | 13106.22 | 13261.07 | 7176.43 | 4486.8 | 2915.65 | 3424.83 | 3351.6 | 3067.75 | 1378.01 | 986.39 | 541.95 | 815.92 |
| 1978 | 8333.03 | 9274.58 | 8199.74 | 4035.12 | 2508.05 | 1693.12 | 2163.39 | 2293.45 | 2205.83 | 882.1 | 690.86 | 837.17 |
| 1979 | 8686.42 | 6269.65 | 6016.26 | 5142.58 | 2807.83 | 1715.91 | 952.79 | 1349.58 | 1449.73 | 1437.71 | 531.28 | 1353.61 |
| 1980 | 13076.4 | 6852.15 | 4288.94 | 3712.31 | 3275.69 | 1770.43 | 1030.27 | 556.59 | 676.95 | 853.96 | 990.7 | 1390.35 |
| 1981 | 33145.83 | 9804.84 | 4816.52 | 2860 | 2430.42 | 2025 | 1192.53 | 651.69 | 300.97 | 376.9 | 558.01 | 2253.4 |
| 1982 | 15680.48 | 26765.62 | 6395.19 | 3247.62 | 1498.27 | 1168.27 | 993.78 | 666 | 359.83 | 163.17 | 192.76 | 3112.94 |
| 1983 | 40831.64 | 12487.92 | 18225.72 | 4336.53 | 1650.93 | 882.84 | 579.18 | 545.8 | 450.27 | 214.98 | 82.92 | 1368.23 |
| 1984 | 26079.33 | 31183.41 | 9226.21 | 10350.72 | 2335.26 | 831.39 | 416.07 | 227.17 | 358.19 | 279.98 | 86.43 | 840.01 |
| 1985 | 22341.19 | 21018.97 | 15516.98 | 5419.27 | 4770.95 | 1120.21 | 433.66 | 194.97 | 138.94 | 234.45 | 175.84 | 690.2 |
| 1986 | 61871.03 | 17183.9 | 13598.58 | 7652.52 | 3367.41 | 2188.74 | 670.13 | 261.85 | 125.24 | 90.23 | 126.8 | 154.27 |
| 1987 | 48649.53 | 49599.77 | 12262.02 | 7086.23 | 2890.31 | 1894.69 | 817.48 | 326.97 | 120.29 | 41.92 | 44.01 | 321.84 |
| 1988 | 44899.26 | 38400.32 | 35367.13 | 6576.48 | 3281.75 | 1470.6 | 1069.87 | 367.99 | 194.41 | 59.57 | 29.79 | 3.91 |
| 1989 | 28604.58 | 35976.81 | 28770.25 | 20310.44 | 2865.3 | 1510.58 | 642.12 | 547.48 | 157.41 | 134.74 | 9.87 | 132.42 |


| YEAR | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 20720.44 | 23011.37 | 24043.49 | 18759.45 | 10171.86 | 1628.37 | 742.71 | 358.33 | 373.13 | 79.11 | 101.27 | 222.62 |
| 1991 | 24974.51 | 16698.44 | 15371.87 | 10528.19 | 7017.24 | 3740.49 | 901.59 | 496.79 | 238.18 | 251.21 | 48.48 | 41.33 |
| 1992 | 19604.36 | 19515.42 | 9034.23 | 5842.58 | 3603.33 | 2590.06 | 1587.57 | 371.7 | 191.38 | 79.19 | 136 | 49.16 |
| 1993 | 23784.03 | 15579.27 | 12297.9 | 4078.56 | 2359.44 | 1707.82 | 1311.64 | 745.13 | 193.03 | 99.69 | 31.35 | 24.86 |
| 1994 | 16884.95 | 18281.95 | 10392.7 | 5825.89 | 1832.69 | 1155.39 | 953.06 | 668.51 | 388.37 | 109.18 | 50.85 | 5.3 |
| 1995 | 38977.69 | 13199.89 | 11383.93 | 6099.24 | 2627.18 | 825.47 | 493.54 | 502.52 | 273.17 | 182.25 | 64.05 | 48.15 |
| 1996 | 24412.3 | 31552.11 | 9885.13 | 6182.4 | 3332.35 | 1085.96 | 363.67 | 186.01 | 237.7 | 129.55 | 83.16 | 22.72 |
| 1997 | 33577.13 | 19718.36 | 24849.12 | 7056.31 | 3750.61 | 1682.31 | 417.69 | 178.31 | 82.62 | 136.7 | 65.35 | 51.17 |
| 1998 | 12772.68 | 27179.37 | 15391.2 | 18136.94 | 4178.37 | 1862.78 | 812.74 | 192.67 | 81.74 | 41.4 | 68.49 | 59.35 |
| 1999 | 58856.59 | 10309.89 | 20724.31 | 10851.3 | 11704.96 | 2173.19 | 907.11 | 332.43 | 88.08 | 37.97 | 8.56 | 62.19 |
| 2000 | 35923.96 | 47896.34 | 7848.36 | 14166.25 | 6576.05 | 5861.61 | 951.33 | 398.84 | 139.16 | 50.39 | 6.66 | 16.51 |
| 2001 | 88189.56 | 28678.22 | 36653.52 | 5082.91 | 7645.11 | 3369.85 | 2333.4 | 464 | 158.25 | 63.27 | 24.97 | 16.57 |
| 2002 | 106023.28 | 71185.56 | 21261.08 | 22375.25 | 2211.61 | 2929.49 | 1366.45 | 703.37 | 114.78 | 48.13 | 30.08 | 3.73 |
| 2003 | 64513.04 | 86531.26 | 50682.08 | 12012.46 | 9464.56 | 1030.75 | 1240.27 | 698.01 | 248.32 | 64.11 | 6.83 | 0 |
| 2004 | 54075.99 | 52520.21 | 68645.24 | 31404.23 | 6221.05 | 3869.89 | 466.59 | 636.32 | 296.41 | 120.97 | 16.3 | 3.19 |
| 2005 | 70045.1 | 44204.91 | 41180.29 | 48471.04 | 17783.42 | 3170.59 | 1533.36 | 142.23 | 255.86 | 110.57 | 8.56 | 0 |
| 2006 | 22264.39 | 56937.28 | 33524.46 | 25132.28 | 24659 | 8136.39 | 1833.08 | 522.49 | 87.49 | 117.19 | 66.1 | 27.39 |
| 2007 | 19344.73 | 16893.91 | 42051.4 | 20406.97 | 13598.46 | 10872.87 | 3259.33 | 921.7 | 172.61 | 42.68 | 85.08 | 28.26 |
| 2008 | 31700.64 | 15086.21 | 10827.5 | 24199.58 | 10850.8 | 7712.29 | 5016.58 | 1276.87 | 387.26 | 68.03 | 24.99 | 0 |
| 2009 | 14067.61 | 21625.54 | 9539.31 | 5609.2 | 11334.62 | 5631.9 | 4302.84 | 2254.12 | 643.67 | 173.19 | 44.84 | 14.9 |
| 2010 | 22829.91 | 11102.26 | 10998.84 | 3305.84 | 2925.72 | 4604.7 | 2793.19 | 1988.26 | 878.25 | 262.78 | 104.7 | 0 |
| 2011 | 33044.78 | 16588.71 | 6451.26 | 4211.26 | 1688.65 | 1482.39 | 1871.66 | 1157.63 | 875.02 | 378.83 | 169 | 19.77 |
| 2012 | 27787.98 | 25338.3 | 11098.81 | 3526.46 | 1815.56 | 950.95 | 734.12 | 895.38 | 476.37 | 319.19 | 185.29 | 42.14 |
| 2013 | 26799.38 | 21973.62 | 11847.98 | 5339.1 | 1754.36 | 671.2 | 503.5 | 323.26 | 372.05 | 182.81 | 138.27 | 126.63 |
| 2014 | 40621.85 | 21289.09 | 13310.66 | 5882.8 | 2342.64 | 973.98 | 360.42 | 301.84 | 177.8 | 172.5 | 72.76 | 4.52 |
| 2015 | 62836.26 | 32463.01 | 15328.09 | 8053.94 | 3295.4 | 1135.31 | 498.83 | 205.51 | 163.88 | 82.23 | 91.47 | 43.84 |
| 2016 | 29626 | 49527.71 | 24535.02 | 10213.26 | 4850.89 | $1916 . .55$ | 673.62 | 246.48 | 90.42 | 71.75 | 37.47 | 69.31 |

Table 6.3.3. Faroe saithe (Division 5.b). Summary table (1961-2015). Values for 2016-2018 are estimates.

| Recruits (age |  |  | Yield (tonnes) | Yield/SSB | Fbar(4-8) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 3) | SSB (TONNES) |  |  |  |
| 1961 | 7827 | 68467 | 9592 | 0,13 | 0.106 |
| 1962 | 12256 | 72862 | 10454 | 0,154 | 0.125 |
| 1963 | 19837 | 76441 | 12693 | 0.173 | 0.114 |
| 1964 | 14811 | 80928 | 21893 | 0,272 | 0.23 |
| 1965 | 22362 | 84690 | 22181 | 0,284 | 0.214 |
| 1966 | 21229 | 87313 | 25563 | 0.3 | 0.25 |
| 1967 | 24897 | 85361 | 21319 | 0.241 | 0.204 |
| 1968 | 22879 | 93938 | 20387 | 0,213 | 0.16 |
| 1969 | 39798 | 103452 | 27437 | 0,274 | 0.191 |
| 1970 | 37092 | 109688 | 29110 | 0,275 | 0.189 |
| 1971 | 38446 | 121970 | 32706 | 0,245 | 0.179 |
| 1972 | 33424 | 137957 | 42663 | 0,308 | 0.236 |
| 1973 | 23621 | 130735 | 57431 | 0.439 | 0.318 |
| 1974 | 19420 | 134010 | 47188 | 0,352 | 0.272 |
| 1975 | 17327 | 135485 | 41576 | 0,307 | 0.297 |
| 1976 | 19709 | 129100 | 33065 | 0.256 | 0.267 |
| 1977 | 13106 | 122228 | 34835 | 0,273 | 0.328 |
| 1978 | 8333 | 105218 | 28138 | 0,266 | 0.243 |
| 1979 | 8686 | 96038 | 27246 | 0.277 | 0.257 |
| 1980 | 13076 | 96219 | 25230 | 0,264 | 0.211 |
| 1981 | 33145 | 85058 | 30103 | 0,37 | 0.382 |
| 1982 | 15680 | 94394 | 30964 | 0,341 | 0.336 |
| 1983 | 40831 | 98647 | 39176 | 0,397 | 0.385 |
| 1984 | 26079 | 104718 | 54665 | 0,522 | 0.478 |
| 1985 | 22341 | 110024 | 44605 | 0,431 | 0.382 |
| 1986 | 61871 | 91607 | 41716 | 0.483 | 0.505 |
| 1987 | 48649 | 94334 | 40020 | 0,441 | 0.396 |
| 1988 | 44899 | 103062 | 45285 | 0,443 | 0.456 |
| 1989 | 28604 | 107481 | 44477 | 0.427 | 0.359 |
| 1990 | 20720 | 103321 | 61628 | 0,608 | 0.561 |
| 1991 | 24974 | 76297 | 54858 | 0,723 | 0.702 |
| 1992 | 19604 | 60153 | 36487 | 0.577 | 0.518 |
| 1993 | 23784 | 59452 | 33543 | 0,555 | 0.45 |
| 1994 | 16884 | 57615 | 33182 | 0,562 | 0.49 |
| 1995 | 38977 | 55735 | 27209 | 0.478 | 0.441 |
| 1996 | 24412 | 60797 | 20029 | 0,319 | 0.343 |
| 1997 | 33577 | 68468 | 22306 | 0,326 | 0.303 |
| 1998 | 12772 | 74278 | 26421 | 0,348 | 0.286 |
| 1999 | 58856 | 77828 | 33207 | 0,419 | 0.334 |
| 2000 | 35923 | 80608 | 39020 | 0,477 | 0.382 |
| 2001 | 88189 | 84237 | 51786 | 0.614 | 0.497 |
| 2002 | 106023 | 81993 | 53546 | 0.653 | 0.479 |
| 2003 | 64513 | 97592 | 46555 | 0,476 | 0.411 |
| 2004 | 54075 | 113454 | 46355 | 0,407 | 0.352 |


| Recruits (age |  |  | Yield (tonnes) | Yield/SSB | Fbar(4-8) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 70045 | 128179 | 67967 | 0.53 | 0.355 |
| 2006 | 22264 | 127839 | 66902 | 0,525 | 0.429 |
| 2007 | 19344 | 121636 | 60785 | 0,501 | 0.394 |
| 2008 | 31700 | 105278 | 57044 | 0.537 | 0.423 |
| 2009 | 14067 | 94514 | 57949 | 0,606 | 0.598 |
| 2010 | 22829 | 70921 | 43885 | 0,618 | 0.551 |
| 2011 | 33044 | 57701 | 29658 | 0.514 | 0.425 |
| 2012 | 27787 | 49796 | 35314 | 0,709 | 0.564 |
| 2013 | 26799 | 46255 | 26463 | 0,572 | 0.447 |
| 2014 | 40621 | 58803 | 23885 | 0.406 | 0.361 |
| 2015 | 62836 | 77216 | 25128 | 0,325 | 0.251 |
| 2016 | 29626 | 96770 | 32085 |  | 0.251 |
| 2017 | 29626 | 126058 | 40403 |  | 0.251 |
| 2018 | 29626 | 144712 |  |  |  |
| Avg. | 31543 | 91844 | 36779 | 0.41 | 0.35 |

Table 6.6.1.1. Faroe saithe (Division 5.b). Input data for prediction with management options for the SPALY assessment.

| 2016 |  | M |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | N | M | MAT | PF | PM | SWT | SEL | CWT |
| 3 | 29626 | 0.2 | 0.04 | 0 | 0 | 1.295 | 0.027 | 1.295 |
| 4 | 49.528 | 0.2 | 0.26 | 0 | 0 | 1.120 | 0.121 | 1.120 |
| 5 | 24535 | 0.2 | 0.51 | 0 | 0 | 1.997 | 0.239 | 1.997 |
| 6 | 10213 | 0.2 | 0.74 | 0 | 0 | 2.719 | 0.311 | 2.719 |
| 7 | 4851 | 0.2 | 0.90 | 0 | 0 | 4.076 | 0.298 | 4.076 |
| 8 | 1917 | 0.2 | 0.98 | 0 | 0 | 5.373 | 0.288 | 5.373 |
| 9 | 674 | 0.2 | 0.98 | 0 | 0 | 5.935 | 0.280 | 5.935 |
| 10 | 246 | 0.2 | 1.00 | 0 | 0 | 6.826 | 0.339 | 6.826 |
| 11 | 90 | 0.2 | 1.00 | 0 | 0 | 6.710 | 0.419 | 6.710 |
| 12 | 72 | 0.2 | 1.00 | 0 | 0 | 7.150 | 1.000 | 7.150 |
| 13 | 37 | 0.2 | 1.00 | 0 | 0 | 7.564 | 1.000 | 7.564 |
| 14 | 69 | 0.2 | 1.00 | 0 | 0 | 9.272 | 1.000 | 9.272 |

2017 (

| AGE | N | M | MAT | PF | PM | SWT | SEL | CWT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 29626 | 0.2 | 0.04 | 0 | 0 | 1.295 | 0.027 | 1.295 |
| 4 | - | 0.2 | 0.26 | 0 | 0 | 1.120 | 0.121 | 1.120 |
| 5 | - | 0.2 | 0.51 | 0 | 0 | 1.997 | 0.239 | 1.997 |
| 6 | - | 0.2 | 0.74 | 0 | 0 | 2.719 | 0.311 | 2.719 |
| 7 | - | 0.2 | 0.90 | 0 | 0 | 4.076 | 0.398 | 4.076 |
| 8 | - | 0.2 | 0.98 | 0 | 0 | 5.373 | 0.288 | 5.373 |
| 9 | - | 0.2 | 0.98 | 0 | 0 | 5.935 | 0.280 | 5.935 |
| 10 | - | 0.2 | 1.00 | 0 | 0 | 6.826 | 0.339 | 6.826 |
| 11 | - | 0.2 | 1.00 | 0 | 0 | 6.710 | 0.419 | 6.710 |
| 12 | - | 0.2 | 1.00 | 0 | 0 | 7.150 | 1.000 | 7.150 |
| 13 | - | 0.2 | 1.00 | 0 | 0 | 7.564 | 1.000 | 7.564 |
| 14 | - | 0.2 | 1.00 | 0 | 0 | 9.272 | 1.000 | 9.272 |


| 2018 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | N | M | MAT | PF | PM | SWT | SEL | CWT |
| 3 | 29626 | 0.2 | 0.04 | 0 | 0 | 1.295 | 0.027 | 1.295 |
| 4 | - | 0.2 | 0.26 | 0 | 0 | 1.120 | 0.121 | 1.120 |
| 5 | - | 0.2 | 0.51 | 0 | 0 | 1.997 | 0.239 | 1.997 |
| 6 | - | 0.2 | 0.74 | 0 | 0 | 2.719 | 0.311 | 2.719 |
| 7 | - | 0.2 | 0.90 | 0 | 0 | 4.076 | 0.298 | 4.076 |
| 8 | - | 0.2 | 0.98 | 0 | 0 | 5.373 | 0.288 | 5.373 |
| 9 | - | 0.2 | 0.98 | 0 | 0 | 5.935 | 0.380 | 5.935 |
| 10 | - | 0.2 | 1.00 | 0 | 0 | 6.826 | 0.339 | 6.826 |
| 11 | - | 0.2 | 1.00 | 0 | 0 | 6.710 | 0.419 | 6.710 |
| 12 | - | 0.2 | 1.00 | 0 | 0 | 7.150 | 1.000 | 7.150 |

[^1]Table 6.6.2.1. Faroe saithe (Division 5.b). Prediction with management option for SPALY assessment.

| 2016 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings |  |  |
| 208397 | 96770 | 1.000 | 0.251 | 32086 |  |  |
| 2017 |  |  |  |  | 2018 |  |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 232095 | 126058 | ก.0000 | 0.0000 | 0 | 288608 | 18.5585 |
|  | 126058 | 0.1000 | 0.0251 | 4520 | 283032 | 180964 |
|  | 126058 | 0.2000 | ก.0.503 | 8927 | 277606 | 176471 |
|  | 126058 | ก.3000 | ก. 0754 | 13209 | 272324 | 172103 |
|  | 126058 | 0.4000 | 0.1006 | 17386 | 267181 | 1678.56 |
|  | 126058 | 0.5000 | 0.1257 | 21455 | 262175 | 163726 |
|  | 126058 | 0.6000 | 0.1508 | 25419 | 257301 | 159709 |
|  | 126058 | 0.7000 | 0.1760 | 29282 | 252554 | 155803 |
|  | 126058 | 0.8000 | 0.2011 | 33047 | 247933 | 152003 |
|  | 126058 | 0.9000 | 0.2263 | 36715 | 243431 | 148308 |
|  | 126058 | 1.0000 | 0.2514 | 40291 | 239048 | 144712 |
|  | 126058 | 1.1000 | 0.2765 | 43776 | 234778 | 141215 |
|  | 126058 | 1.2000 | 0.3017 | 47174 | 230619 | 137813 |
|  | 126058 | 1.3000 | 0.3268 | 50486 | 226568 | 134503 |
|  | 126058 | 1.4000 | 0.3520 | 53715 | 222622 | 131282 |
|  | 126058 | 1.5000 | 0.3771 | 56864 | 218777 | 128148 |
|  | 126058 | 1.6000 | 0.4022 | 59934 | 215031 | 125099 |
|  | 126058 | 17000 | 0.4274 | 62928 | 211381 | 122131 |
|  | 126058 | 18000 | 0.4525 | 65848 | 207825 | 119243 |

Input units are thousands and kg - output in tonnes

Table 6.7.1.1. Faroe saithe (Division 5.b). Yield-per-recruit input data.

| AGE | M | MAT |  | PF |  | PM | WEST | SEL |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 0.2 | 0.02 | 0 | 0 | 1.304 | 0.048 | WECA |  |
| 4 | 0.2 | 0.21 | 0 | 0 | 1.304 |  |  |  |
| 5 | 0.2 | 0.47 | 0 | 0 | 2.031 | 0.278 | 1.668 |  |
| 6 | 0.2 | 0.71 | 0 | 0 | 2.602 | 0.5118 | 2.031 |  |
| 7 | 0.2 | 0.86 | 0 | 0 | 3.373 | 0.52 | 3.373 |  |
| 8 | 0.2 | 0.95 | 0 | 0 | 4.318 | 0.5648 | 4.318 |  |
| 9 | 0.2 | 0.99 | 0 | 0 | 5.085 | 0.5572 | 5.085 |  |
| 10 | 0.2 | 1 | 0 | 0 | 5.904 | 0.6514 | 5.904 |  |
| 11 | 0.2 | 1 | 0 | 0 | 6.777 | 0.7174 | 6.777 |  |
| 12 | 0.2 | 1 | 0 | 0 | 7.472 | 0.5888 | 7.472 |  |
| 13 | 0.2 | 1 | 0 | 0 | 7.835 | 0.4844 | 7.835 |  |
| 14 | 0.2 | 1 | 0 | 0 | 9.388 | 0.4844 | 9.388 |  |

Table 6.9.1. Faroe saithe (Division 5.b). Comparison between the current assessment (NWWG2016 SPALY) statistical assessment (NWWG2016 ADMB) and predictions from last year in the terminal year (2015).

|  | NWWG2015 <br> PREDICTION |  |  |  | NWWG2016 <br> (SPALY) | NWWG2016 (ADMB) |
| :--- | :--- | :--- | :--- | :---: | :---: | :---: |
| Recruitment | 27 mill. | 62 mill. | 36 mill. |  |  |  |
| SSB | $82089 \mathrm{t}$. | 77000 t. | 68278 t. |  |  |  |
| Fbar(4-8) | 0.310 | 0.25 | 0.32 |  |  |  |
| Landings | 35360 t. | 25128 t. | 26482 t. |  |  |  |

### 6.18 Figures



Figure 6.2.1.1. Faroe saithe (Division 5.b). Landings in 1000 tonnes (1961-2015). Horizontal red line represents historical average landings.


Figure 6.2.1.2. Saithe in the Faroes (Division 5.b). Cumulative domestic landings (2000-2016).


Figure 6.2.3.1. Faroe saithe (Division 5.b). Mean weight at age (kg) in commercial catches (ages 3-9) (1961-2018). Weights from 2016 to 2018 are estimates. Horizontal lines show historical average.


Figure 6.2.4.1. Faroe saithe (Division 5.b). Smoothed maturity ogives (ages 3-8)(1983-2015) from FGFS1 (spring survey). Horizontal lines show historical average.


Figure 6.2.5.1.1. Faroe saithe (Division 5.b). Predicted catch rates from the commercial fleet (pairtrawlers) used for tuning the assessment (black line). Catch rates (kg/hour) from the Faroese bottom-trawl fall FGFS2 (1996-2015)(red line) and spring survey FGFS1 (1994-2016)(blue line). Shade areas show standard errors in the estimation of indices.


Figure 6.2.5.1.2. Faroe saithe (Division 5.b). Length composition from the Faroese bottom-trawl spring survey FGFS1 (1994-2016)


Figure 6.2.5.1.3. Faroe saithe (Division 5.b). Length composition from the Faroese bottom-trawl summer survey FGFS2 (1996-2015)


Figure 6.2.5.1.4. Faroe saithe (Division 5.b). Age-disaggregated indices in the Faroese bottom-trawl spring survey FGFS1 (ages 3-10, years 1994-2016)


Figure 6.2.5.1.5. Faroe saithe (Division 5.b). Age-disaggregated indices in the Faroese bottom-trawl fall survey FGFS2 (ages 3-10, years 1996-2015)


Figure 6.2.5.1.6. Faroe saithe (Division 5.b). Numbers from spring survey (FGFS1) plotted against numbers of the same year class one year later. Letters in the figures represent year classes.


Figure 6.2.5.1.7. Faroe saithe (Division 5.b). Numbers from summer survey (FGFS2) plotted against numbers of the same year class one year later. Letters in the figures represent year classes.


Figure 6.2.5.2.1. Faroe saithe (Division 5.b). Age-disaggregated indices in the commercial pair-trawl fleet (ages 3-10, years 1995-2015)


Figure 6.2.5.2.2. Faroe saithe (Division 5.b). Indices from in the commercial pair-trawl plotted against indices of the same year class one year later. Letters in the figures represent year classes.


Figure 6.3.1. Faroe saithe (Division 5.b). Log-catchability residuals of the spaly assessment calibrated with the commercial series (ages 3-11, years 1995-2015). Blue and red bubbles represent positive and negative residuals respectively.
catch residuals


Figure 6.3.3. Faroe saithe (Division 5.b). Catch-(ages 3-14+, years 1961-2015)(top plot) and survey-at-age (ages 3-11, years 1995-2015)(bottom plot) residuals from a statistical catch-at-age model. Red and white bubbles represent positive and negative residuals respectively.


Figure 6.4.1.1. Faroe saithe (Division 5.b). EqSim simulation. Stock-recruitment function used in the simulations (Hockey-stick).


Figure 6.4.1.2. Faroe saithe (Division 5.b). EqSim simulation outputs with assessment errors and Hockey-stick function from WKMSYREF2 report. $B_{\text {lim }}$ is undefined but was set as $B_{\text {lim }}=B_{\mathrm{pa}} / 1.4$.


Figure 6.4.1.3. Faroe saithe (Division 5.b). Stock-recruitment plot in relation to $\mathrm{F}_{\text {low }}=0.13$ (lowest regression line), $\mathrm{F}_{\text {med }}=0.31$ (middle regression line) and Fhigh $=0.79$ (top regression line). Vertical red line represents $B_{\text {trigger }}=55000 t$.


Figure 6.5.1. Faroe saithe (Division 5.b). Recruitment (age 3) in millions (top-left), total-stock biomass (thousand tonnes)(top-middle), spawning-stock biomass (thousand tonnes) (bottom-left), landings (thousand tonnes)(middle-left), landings SSB ratio (middle-middle), Fbar (ages 4 to 8)(mid-dle-right), reference biomass (B4+) (thousand tonnes) (bottom-left) and landings B4+ ratio (bottomright). Black line represents the spaly run. Red lines show estimates from a catch-at-age statistical model implemented in ADMB. Horizontal blue lines represent historical averages.


Figure 6.5.2. Faroe saithe (Division 5.b). Fishing mortality (average over ages 4-8)(1961-2015)


Figure 6.5.3. Faroe saithe (Division 5.b). Recruitment-at-age 3 (millions)(1961-2016). The 2016 recruitment estimate is used in the short-term forecast.


Figure 6.5.4. Faroe saithe (Division 5.b). Spawning-stock biomass ('000 tonnes)(1961-2016). The 2016 SSB estimate is used in the short-term forecast. Horizontal lines represent $B_{\text {trigger }}=B_{\mathrm{pa}}=55000 \mathrm{t}$.


Figure 6.5.6. Faroe saithe (Division 5.b). Numbers of mature fish in the stock (ages 3-9) for 2006, 2013 and 2014.


Figure 6.5.7. Faroe saithe (Division 5.b). SSB Recruitment (age 3) plot. $B_{\text {trigger }}=B_{p a}=55000 \mathrm{t}$.


Figure 6.6.1.1. Faroe saithe (Division 5.b). Residual plots from a 3-year running average weight model and the model in which weights are predicted from the previous year in the same year class. Red and white bubbles represent positive and negative residuals respectively.


Figure 6.6.1.2. Faroe saithe (Division 5.b). Observed (stapled lines) and predicted weights (solid lines)(ages 4-8, years 1985-2015)


Figure 6.6.2.1. Faroe saithe (Division 5.b). Short-term prediction output (spaly assessment). Solid and broken lines represent landings ( $\mathbf{t}$ ) and spawning-stock biomass ( $\mathbf{t}$ ) respectively.


Figure 6.6.2.2 Faroe saithe (Division 5.b). Composition of landings (upper figure) and SSB (lower figure) by year classes in 2016.


Figure 6.7.1.1. Faroe saithe (Division 5.b). Yield and spawning per-recruit calculations. Dashed and solid lines represent Yield/R and SSB/R respectively.


Figure 6.8.1. Faroe saithe (Division 5.b). Retrospective analysis of recruitment-at-age 3 (millions)(top figure), spawning-stock biomass ('000 tonnes)(middle figure) and average fishing mortality over age groups 4-8 (bottom figure) from the spaly assessment.

## 7 Overview on ecosystem, fisheries and their management in Icelandic waters

This section gives a very broad and general overview of the marine ecosystem, fishery, fleet, species composition and some bycatch analysis of the commercially landed species as well as management measures in the Icelandic Exclusive Economic Zone. The Icelandic EEZ covers partly the IIa2, Va1, Va2, Vb1b, XIIa4, XIVa and XIVb2 ICES statistical regions. In practice however, the Icelandic landings of different species are generally reported as catches/landings in Va.
The information on the ecosystem of Icelandic waters is brief but a more detailed description is available in the WGRED report (ICES 2008).

### 7.1 Environmental and ecosystem information

Iceland is located at the junction of the Mid-Atlantic Ridge and the Greenland-Scotland Ridge just south of the Arctic Circle and this is reflected in the topography around the country. Substrate characteristics can be largely influenced by depth. Hard bottom is more often found in shallower waters compared to deep waters. In deeper waters, hard bottom is often confined to abrupt features such as ridges and seamounts. Soft sediments often dominate in the troughs and outside the continental slope. The shelf around Iceland is narrowest off the south coast (Figure 7.3.4) and is cut by submarine canyons around the country (Figure 7.3.4).

The Polar Front lies west and north off Iceland and separates the cold and southward flowing waters of Polar origin from the northward flowing waters of Atlantic origin. South and east off Iceland the North Atlantic Current flows towards the Norwegian Sea. The Irminger Current is a branch of the North Atlantic Current and flows northwards over and along the Reykjanes Ridge and along the western shelf break. In the Denmark Strait it divides into a branch that flows northeastward and eastward to the waters north off Iceland, as the North Icelandic Irminger Current, and another branch that flows south-westward along the East Greenland Current. In the Iceland Sea north off Iceland, a branch originating from the cold East Greenland Current flows over the Kolbeinsey Ridge and continues to the southeast along the northeastern shelf brake as the East Icelandic Current, which is part of a cyclonic gyre in the Iceland Sea. This current subsequently continues into the Norwegian Sea along the Atlantic water flowing eastwards over the Iceland-Faroes Ridge (Stefansson 1962, Valdimarsson and Malmberg 1999).

The Icelandic Shelf is a high ( $150-300 \mathrm{gC} / \mathrm{m} 2-\mathrm{yr}$ ) productivity ecosystem according to SeaWiFS global primary productivity estimates. Productivity is higher in the southwest regions than to the northeast and higher on the shelf areas than in the oceanic regions (Gudmundsson 1998). In terms of abundance, copepods dominate the mesozooplankton within Icelandic waters with Calanus finmarchicus being the most abundant species, often comprising between $60-80 \%$ of net-caught zooplankton in the uppermost 50 m (Astthorsson and Vilhjalmsson 2002, Astthorsson et al. 2007).

The structure of benthic communities in Icelandic waters is likely to be influenced by a large number of factors. Amongst these, water mass characteristics will have profound effects on species composition and spatial distribution patterns at the largest spatial scales (e.g. $>50 \mathrm{~km}$ ) whereas substrate characteristics (e.g. sediment type and rugosity) and topography will have profound effects on smaller scales (e.g. meters to
kilometers), (e.g. Weisshappel and Svavarsson 1998). Shrimp biomass in Icelandic waters, both in inshore and offshore waters, has been declining in recent years. Consequently the fishing effort was reduced and is now banned in most inshore areas. The causes for the decline in the inshore shrimp biomass is in part considered to be environmentally driven, both due to increasing water temperature north of Iceland and due to increasing biomass of younger cod, haddock and whiting.

Based on information from fishermen, eleven cold-water coral areas were known to exist close to the shelf break off the northwest towards southeast Iceland around 1970. During the 70 s and 80 s, more coral areas were found by fishermen as a direct consequence of the bottom trawling fisheries extending into deeper waters. More recently there has been a considerable effort in mapping cold-water coral habitats in Icelandic waters and to investigate their biology using the state of the art technology such as unmanned submersibles. At present, large cold-water coral areas have been located on the Reykjanes Ridge and on the shelf break south and southeast Iceland (Steingrímsson and Einarsson 2004). Many of the cold-water coral areas that have been surveyed have already been destroyed. Currently, 5 areas with relatively undisturbed cold-water corals have received full protection and several other areas are under consideration for further protection.

The database of the BIOICE programme provides information on the spatial distribution of benthic organisms within the Icelandic territorial waters based on samples collected from 579 locations, including horny corals (Gorgonacea) and seapens (Pennatulacea) that are considered sensitive to fishing. Gorgonian corals occur all around Iceland but these are relatively uncommon on the shelf ( $<500 \mathrm{~m}$ depth) but can be found in relatively high numbers in deep waters ( $>500 \mathrm{~m}$ ) off south, west and north coasts of Iceland, given the right environmental conditions. Similar distribution patterns were observed in the distribution of pennatulaceans, these being common in deeper waters, especially off South Iceland (Guijarro et al. 2007).

About 25 species of stocks of fish and marine invertebrates are exploited commercially on a regular basis in Icelandic waters.

Icelandic waters are comparatively rich in species and contain around 30 commercially exploited stocks of fish and marine invertebrates. The most important commercial species are cod, haddock, saithe, redfish, Greenland halibut and various other flatfish, wolffish, tusk (Brosme brosme), ling (Molva molva), herring, capelin and blue whiting. Most fish species spawn in the warm Atlantic water off the south and southwest coasts. Fish larvae and 0 -group subsequently drift west and then north from the spawning grounds to nursery areas on the shelf off northwest, north and east Iceland, where they grow in a mixture of Atlantic and Arctic water.

Capelin is important in the diet of cod as well as a number of other fish stocks, marine mammals and seabirds. Unlike other commercial stocks, adult capelin undertake extensive feeding migrations north into the cold waters of the Denmark Strait and Iceland Sea during summer. Capelin abundance has been oscillating on roughly a decadal period since the 1970s, producing a yield of up to 1600 Kt at the most recent peak. In recent years the stock size of capelin has decreased from about 2000 Kt in 1996/97 to about 900 Kt in 2012/13 (Anon. 2013). Herring were very abundant in the early 1960s until the stock collapsed in the nineteen sixties due to overfishing. From 1970 onwards the stock size has increased until attaining historical high levels in the last decade. Abundance of demersal species have been generally trending downward since the 1950s with total catches dropping from over 800 Kt to less than 500 Kt in the early 2000s.

A number of species of sharks and skates are known to be caught as a by-catch in Icelandic waters, but information on amount of the catches is incomplete, and the status of these species is not known. Information on status and trends of non-commercial species are collected in extensive bottom trawl surveys conducted in early spring and autumn.

The seabird community in Icelandic waters is composed of relatively few but mostly abundant species, accounting for roughly $1 / 4$ of total number and biomass of seabirds within the whole ICES area (ICES 2002). Auks and petrels are the most important groups, comprising almost $3 / 5$ and $1 / 4$ of the total abundance and biomass in the area, respectively. The estimated annual food consumption is on the order of 1.5 million tonnes.

At least 12 species of cetaceans occur regularly in Icelandic waters, and additional 10 species have been recorded more sporadically. In the continental shelf area, the minke whale (Balaenoptera acutorostrata) probably has the largest biomass. Based on the 2001 sightings survey, 67000 minke whales were estimated in the Central North Atlantic stock region, with 44000 animals in Icelandic coastal waters (NAMMCO 2004, Borchers et al. 2003, Gunnlaugsson 2003). In the 2007 aerial survey the abundance of minke whales was estimated at around 21000 animals on the Icelandic shelf. The reasons for this decrease are not known. Two species of seals, common seal (Phoca vitulina) and grey seal (Halichoerus grypus) breed in Icelandic waters, while 5 other species are found as vagrants (Sigurjonsson and Hauksson, 1994; Hauksson, 1993, 2004).

### 7.2 Environmental drivers of productivity

Mean weight at age of Icelandic cod have been shown to correlate well with the size of the capelin stock and therefore the capelin stock was used as a predictor of weights in the landings in 1991 - 2007. In 1981 - 1982, cod weights were low following collapse of the capelin stock and were also relatively low in 1990-1991 when the capelin stock was small. In recent years this relationship seems to be much weaker and have not been used for predictions. The reasons for these changes are most likely changes in the spatial distribution of capelin or uncertainties in the estimation of the capelin stock size.
No other ecosystem drivers of productivity that may affect the assessment of the Icelandic stocks assessed in this report were presented to the NWWG in 2013.

### 7.3 Ecosystem considerations (General)

After 1996 a rise in both temperature and salinity were observed in the Atlantic water south and west of Iceland. Temperature and salinity have remained at similar high levels since and west of Iceland amounts to an increase of temperature of about $1^{\circ} \mathrm{C}$ and salinity by one unit on average (Figure 7.3.1.) and these changes can therefore be regarded as conspicuous. Off central N-Iceland, similar trends have been observed although with higher inter-annual variability. This period has been characterized with an increase of temperature and salinity in the winter north of Iceland in the last $12-14$ years which is on average above $1^{\circ} \mathrm{C}$ and 1 salinity units. (Figure 7.3.2)

It appears that these changes in seawater temperature have had considerable effects on the spatial distribution of fish species in Icelandic waters with many species now found further northwards. The most obvious examples of such changes is the increased abundance of haddock, mackerel, whiting, monkfish, lemon sole and witch in the mixed water area north of Iceland.

On the other hand, coldwater species like Greenland halibut and northern shrimp have become scarcer. Capelin have shifted their larval drift and nursing areas westwards to the colder waters off E-Greenland. Furthermore, the arrival of adult capelin to the overwintering grounds on the outer shelf off N -Iceland has been delayed and migration routes to the spawning grounds off S- and W-Iceland are currently located farther off N - and E-Iceland and do not reach as far west along the south coast as was the rule in most earlier years (Figure 7.3.3. and 7.3.4.). These changes in the spatial distribution patterns of capelin may have had an effect on the growth rate of various predators, as is reflected in low weight of cod in recent years.

There is one demersal stock, which apparently has not taken advantage, or not been able to take advantage, of the milder marine climate of Icelandic waters. This is the Icelandic cod, which was very abundant during the last warm epoch, which began around 1920 and lasted until 1965. By the early 1980s the cod stock had been fished down to much lower levels as compared to previous decades and has remained relatively low since. During the last 20 years the Icelandic cod stock has not produced a large year class and the average number of age 3 recruits in the last 20 years is about 150 million fish per annum, as compared to $205-210$ recruits in almost any period prior to that, even during the ice years of 1965-1971. Immigrants from Greenland are not included in this comparison. It is not possible to pinpoint exactly what has caused this change, but a very small and young spawning stock is the most obvious common denominator for this protracted period of impaired recruitment to the Icelandic cod stock. Regulations, particularly the implementation of the catch rule in 1993 have resulted in lower fishing mortalities in the last ten years when compared with the years prior to 2000. Further, despite the overall low recruitment, this reduction in fishing mortality has almost resulted in almost doubling of the spawning stock biomass. This increase in the SSB biomass has however not resulted in significant increase in recruitment in recent years, although year classes 2008 and 2009 are now estimated around average size.

Associated with the large warming of the 1920s, was a well documented drift of larval and 0-group cod as well as some other fish species, from Iceland across the northern Irminger Sea to East and then West-Greenland. Although many of these fish apparently returned to Iceland to spawn and did not leave again, there is little doubt that the cod, remaining in West-Greenland waters which also had warmed, were instrumental in establishing a self-sustaining Greenlandic cod stock that eventually became very large. It seems that significant numbers of cod of the 2003 year class have drifted across to Greenland in that year. Tag returns, survey estimates in Greenlandic waters as well as anomalies in the catch-at-age matrix in Iceland indicate that a portion of the moderate 2003 year class that has been observed in Greenlandic waters in recent years may have migrated to Icelandic waters in 2009.

### 7.4 Description of fisheries [Fleets]

Only Icelandic vessels are considered in the following analysis since they constitute the largest operational players in Icelandic waters. Few trawlers and longliners of other nationalities operate in the Icelandic region principally targeting deep-sea redfish, cod, tusk, ling and, with some bycatch of other species. Additionally some limited pelagic fishery of foreign boats on capelin, herring and blue whiting also takes place in Icelandic waters.

The data sources used in this section are landings, boat, log book and discard databases. Landings of species by each boat and gear are effectively available electronically
in real time (end of day of landing). Log-book statistics are generally available in a centralized database about 1 month after the day of fishing operation. Since 2009 increasing proportion of vessels are using electronic logbooks. Fisheries scientists have direct access to the logbook database.
The Icelandic fishing fleet can be characterised by the most sophisticated technological equipment available in this field. This applies to navigational techniques and fish-detection instruments as well as the development of more effective fishing gear. The most significant development in recent years is the increasing size of pelagic trawls and with increasing engine power the ability to catch pelagic fishes at greater depths than previously possible. There have also been substantial improvements in recent decades with respect to technological aspects of other gears such as bottom trawl, longline and handline. Each fishery uses a variety of gears and some vessels frequently shift from one gear to another within each year. The most common demersal fishing gear are otter trawls, longlines, seines, gillnets and jiggers while the pelagic fisheries use pelagic trawls and purse seines. The total recorded landings of the Icelandic fleets in 2010 amounted to around 1 million tonnes where pelagic fishes amounted to 0.5 million tonnes. Spatial distributions of the catches are shown in figure 7.4.1. Detailed information of landings by species and gear type are given in Table 7.1. Spatial overviews of the removal of the some important species by different gear are given in Figures 7.4.2.-7.4.5.

A simple categorization of boats among the different fisheries types is impossible as many change gear depending on fish availability in relation to season, quota status of the individual companies, fish availability both in nature and on the quota exchange market, market price, etc. E.g. larger trawl vessels may operate both on demersal species using bottom trawls as well as using purse seine and pelagic trawls on pelagic species. Total number of vessels within each fleet category in 2010 is thus limited to the broad categories given below:

| Type | No. vessels 1) | Gear type used |
| :--- | :--- | :--- |
| Trawlers | 57 | Pelagic and bottom trawl |
| Vessels $>100 \mathrm{t}$ | 140 | Purse seine, longline, trawl, gillnet |
| Vessels $<100 \mathrm{t}$ | 621 | Gillnet, longline, danish seine, trawl, jiggers |
| Open boats | 807 | Jiggers, longliners (including recreational fishers) |
| Total | 1625 |  |
| 1)Source: Statistic Iceland $-\mathrm{http}: / /$ www.statice.is/ |  |  |

The demersal fisheries take place all around Iceland including variety of gears and boats of all sizes. The most important fleets targeting them are:

Large and small trawlers using demersal trawl. This fleet is the most important one fishing cod, haddock, saithe, redfish as well as a number of other species. This fleet is operating year around; mostly outside 12 nautical miles from the shore.

Boats (<300 GRT) using gillnet. These boats are mostly targeting cod but haddock and a number of other species are also target. This fleet is mostly operating close to the shore.

Boats using longlines. These boats are both small boats ( $<10$ GRT) operating in shallow waters as well as much larger vessels operating in deeper waters. Cod and haddock are the main target species of this fleet but a number other species are also caught, some of them in directed fisheries.

Boats using jiggers. These are small boats ( $<10$ GRT). Cod is the most important target species of this fleet with saithe of secondary importance.

Boats using Danish seine. (20-300GRT) Cod, haddock and variety of flatfishes, e.g. plaice, dab, lemon sole and witch are the target species of this fleet.

Although different fleets may be targeting the main species the spatial distribution of effort may different. In general it can be observed that the bottom trawl fleet is fishing in deeper waters than the long line fleet (Figures 7.4.6. and 7.4.7).

The pelagic fisheries targeting capelin, herring, blue whiting and mackerel is almost exclusively carried out by larger vessels. The fisheries in Icelandic waters for capelin and herring are carried out using both purse seine and pelagic trawl while that of blue whiting and mackerel is exclusively carried out with pelagic trawl. Additionally a significant part of the pelagic fisheries of the Icelandic fleet is caught outside the Icelandic EEZ, both on the Atlanto-Scandian herring and on blue whiting.

### 7.5 Regulations

The Ministry of Fisheries is responsible for management of the Icelandic fisheries and implementation of the legislation. The Ministry issues regulations for commercial fishing for each fishing year, including an allocation of the TAC for each of the stocks subject to such limitations. Below is a short account of the main feature of the management system.

### 7.5.1 The ITQ system

A system of transferable boat quotas was introduced in 1984. The agreed quotas were based on the Marine Research Institute's TAC recommendations, taking some socioeconomic effects into account. Until 1990, the quota year corresponded to the calendar year but since then the quota, or fishing year, starts on September 1 and ends on August 31 the following year. This was done to meet the needs of the fishing industry. In 1990, an individual transferable quota (ITQ) system was established for the fisheries and they were subject to vessel catch quotas. Since 2006/2007 fishing season, all boats operate under the TAC system.

With some minor exceptions it is required by law to land all catches. Consequently, no minimum landing size is in force. To prevent fishing of small fish various measures such as mesh size regulation and closure of fishing areas are in place (see below).

Within this system individual boat owners have substantial flexibility in exchanging quota, both among vessels within individual company as well as among different companies. The latter can be done via temporary or permanent transfer of quota. In addition, some flexibility is allowed by individual boats with regard to transfer allowable catch of one species to another. These measures, which can be acted on more or less instantaneously, are likely to result in lesser initiative to discards and misreporting than can be expected if individual boats are restricted by strict TAC measures alone. They may however result in fishing pressures of individual species to be different than intended under the single species TAC allocation.

### 7.5.2 Mesh size regulations

With the extension of the fisheries jurisdiction to 200 miles in 1975, Iceland introduced new measures to protect juvenile fish. The mesh size in trawls was increased from 120 mm to 155 mm in 1977. Mesh size of 135 mm was only allowed in the fisheries for redfish in certain areas. Since 1998 a minimum mesh size of 135 is allowed in the
codend in all trawl fisheries not using "Polish cover" and in the Danish seine fisheries. For the gillnet fishery both minimum and maximum mesh-sizes are restricted. Since autumn 2004 the maximum allowed mesh-size in the gillnet fishery is 8 inches. The objective of this measure is to decrease the effort directed towards bigger spawners.

### 7.5.3 Area closures

Real time area closure: A quick closure system has been in force since 1976 with the objective to protect juvenile fish. Fishing is prohibited for at least two weeks in areas where the number of small fish in the catches has been observed by inspectors to exceed certain percentage ( $25 \%$ or more of $<55 \mathrm{~cm}$ cod and saithe, $25 \%$ or more of $<45 \mathrm{~cm}$ haddock and $20 \%$ or more of $<33 \mathrm{~cm}$ redfish). If, in a given area, there are several consecutive quick closures the Minister of Fisheries can with regulations close the area for longer time forcing the fleet to operate in other areas. Inspectors from the Directorate of Fisheries supervise these closures in collaboration with the Marine Research Institute. In 2010, 113 such closures took place:

Permanent area closures: In addition to allocating quotas on each species, there are other measures in place to protect fish stocks. Based on knowledge on the biology of various stocks, many areas have been closed temporarily or permanently aiming at protect juveniles. Figure 7.5 .1 shows map of such legislation that was in force in 2004. Some of them are temporarily, but others have been closed for fishery for decades.

Temporary area closures: The major spawning grounds of cod, plaice and wolfish are closed during the main spawning period of these species. The general objectives of these measures, which were in part initiated by the fishermen, are to reduce fishing during the spawning activity of these species.

### 7.5.4 Discards

Discarding measurements have been carried out in Icelandic fisheries since 2001, based on extensive data collection and length based analysis of the data (Pálsson 2003). The data collection is mainly directed towards main fisheries for cod (Gadus morhua) and haddock (Melanogrammus aeglefinus) and towards saithe (Pollachius virens) and golden redfish (Sebastes marinus) fisheries in demersal trawl and plaice in Danish seine. Sampling for other species is not sufficient to warrant a satisfactory estimation of discarding. The discard rate for cod has been in the range of $0.2-2.2 \%$ of the reported landings over the time investigated (Figure 7.5.2.). The discard estimates for haddock are somewhat higher ranging between $0.7-5 \%$ annually. Discarding of saithe and golden redfish has been negligible over time period of investigation. Estimates of discards of cod and haddock in 2010 by individual fleets are given in table 7.2. These relatively low discard rates compared to what is generally assumed to be a side effect of a TAC system may be a result of the various measures, including the flexibility within the Icelandic ITQ system (see above). Since the time series of discards is relatively short it is not included in the assessments.

All catch that is brought ashore must by law be weighted by a licensed body. The monitoring and enforcement is under the realm of the Directorate of Fisheries. Under the TAC system there are known incentives for misreporting, both with regards to the actual landings statistics as well as with regards to the species recorded. This results in bias in the landings data but detailed quantitative estimates of how large the bias may be, is not available to the NWWG. Unpublished report from the Directorate of Fisheries, partly based on investigation comparing export from fish processing plants with
the amount of fish weighted in the landing process indicate that this bias may be of the order of single digit percentages and not in double digits.

### 7.6 Mixed fisheries, capacity and effort

A number of species caught in Icelandic waters are caught in fisheries targeting only one species, with very little bycatch. These include the pelagic fisheries on herring, capelin and blue whiting (see however below), the Greenland halibut fishery in the west and southeast of Iceland and the S. mentella fishery. Advice given for these stocks should thus not influence the advice of other stocks.

Other fisheries, particularly demersal fisheries may be classified as more mixed, where a target species of e.g. cod, haddock, saithe or S. marinus may be caught in a mixture with other species in the same haul/setting (Figure 7.6.1.). Fishermen can however have a relatively good control of the relative catch composition of the different species. E.g. the saithe fishery along the shelf edge is often in the same areas as the redfish fisheries: Fleets are often targeting at redfish during daytime and saithe during nights. Therefore the fishery for one of those species is relatively free of bycatch of the other species even though they take place in the same area. Small differences in the location of setting are also known to affect the catch composition. This has for example been documented in the long line fisheries in Faxabay, where in adjacent areas cod catches and wolfish catches are known to consistently dominate the catches in individual setting. There are however numerous species in Icelandic waters that can be classified as "bycatch species" in some fisheries. E.g. in the bottom trawl fisheries $75 \%$ of the annual plaice yield is caught in hauls where plaice is minority of the catches. In a proper fisheries based advice taking mixed fisheries issues into account, such stocks may have a greater influence on the advice on the main stocks that are currently assessed by ICES than fisheries linkage among the latter.
In the pelagic fisheries catch other than the targeted species is considered rare. In some cases juveniles of other species are caught in significant numbers. When observers are on board or when fishermen themselves provide voluntary information, the fishing areas have in such cases been closed for fishing, temporarily or permanently. By catch of adults of other species in the blue whiting fishery have been estimated (Pálsson 2005).

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Table 7.1 Overview of the 2010 landings of fish and marine invertebrates caught by the Icelandic fleet categorized by gear types. Based on landing statistics from the Directorate of Fisheries. Landings are given in thous. tonnes.

| Species/gear | Long line | Gillnets Jiggers |  | Danish seine | Bottom trawl | Nephros trawl | Pelagic trawl | Purse seine | Shrimp trawl | redge | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Herring | 0.000 | 0.000 | 0.000 | 0.000 | 0.112 | 0.000 | 213.528 | 40.836 | 0.000 | 0.000 | 0.000 | 254.476 |
| Cod | 57.493 | 16.552 | 3.721 | 8.285 | 82.996 | 1.581 | 0.923 | 0.009 | 1.006 | 0.000 | 0.784 | 173.349 |
| Mackerel | 0.000 | 0.001 | 0.180 | 0.000 | 0.164 | 0.000 | 121.680 | 0.001 | 0.000 | 0.000 | 0.000 | 122.028 |
| Capelin | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 3.187 | 112.328 | 0.000 | 0.000 | 0.000 | 115.515 |
| Blue whiting | 0.000 | 0.000 | 0.000 | 0.000 | 0.124 | 0.000 | 87.784 | 0.000 | 0.000 | 0.000 | 0.000 | 87.908 |
| Haddock | 23.916 | 0.380 | 0.012 | 10.137 | 29.481 | 0.212 | 0.630 | 0.000 | 0.041 | 0.000 | 0.028 | 64.836 |
| Saithe | 0.594 | 4.453 | 2.383 | 1.093 | 42.441 | 0.404 | 1.216 | 0.000 | 0.007 | 0.000 | 0.068 | 52.660 |
| Golden redfish | 1.080 | 0.194 | 0.058 | 0.513 | 35.777 | 0.932 | 0.594 | 0.000 | 0.014 | 0.000 | 0.014 | 39.176 |
| Pearlside | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 17.912 | 0.000 | 0.000 | 0.000 | 0.000 | 17.912 |
| Atlantic argentine | 0.000 | 0.000 | 0.000 | 0.000 | 16.321 | 0.001 | 0.256 | 0.000 | 0.000 | 0.000 | 0.000 | 16.579 |
| Golden redfish | 0.000 | 0.000 | 0.000 | 0.000 | 1.921 | 0.000 | 12.872 | 0.000 | 0.000 | 0.000 | 0.000 | 14.794 |
| Deepwater redfish | 0.052 | 0.002 | 0.000 | 0.000 | 14.149 | 0.000 | 0.181 | 0.000 | 0.000 | 0.000 | 0.000 | 14.384 |
| Greenland halibut | 0.033 | 0.000 | 0.000 | 0.000 | 12.147 | 0.000 | 0.263 | 0.000 | 0.861 | 0.000 | 0.001 | 13.305 |
| Atlantic catfish | 6.915 | 0.020 | 0.002 | 1.032 | 4.490 | 0.083 | 0.033 | 0.000 | 0.000 | 0.000 | 0.027 | 12.602 |
| Ling | 6.529 | 0.363 | 0.011 | 0.404 | 1.538 | 0.981 | 0.011 | 0.000 | 0.000 | 0.000 | 0.028 | 9.865 |
| Shrimp | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.155 | 0.000 | 7.607 | 0.000 | 0.000 | 7.762 |
| Tusk | 6.760 | 0.052 | 0.003 | 0.000 | 0.093 | 0.005 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 6.915 |
| Blue Lling | 3.978 | 0.091 | 0.000 | 0.092 | 1.901 | 0.283 | 0.013 | 0.000 | 0.002 | 0.000 | 0.015 | 6.375 |
| Plaice | 0.105 | 0.118 | 0.006 | 3.640 | 2.020 | 0.003 | 0.015 | 0.000 | 0.001 | 0.000 | 0.077 | 5.984 |
| Monkfish | 0.079 | 0.176 | 0.001 | 0.430 | 0.452 | 0.556 | 0.000 | 0.000 | 0.001 | 0.000 | 1.586 | 3.281 |
| Whiting | 0.425 | 0.030 | 0.002 | 0.191 | 2.037 | 0.155 | 0.000 | 0.000 | 0.001 | 0.000 | 0.001 | 2.842 |
| Redfish | 0.001 | 0.000 | 0.000 | 0.000 | 2.446 | 0.000 | 0.154 | 0.000 | 0.000 | 0.000 | 0.000 | 2.601 |
| Nephrops | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.541 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.541 |
| Sea cucumber | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.246 | 0.000 | 2.246 |
| Lumpfish roe | 0.000 | 0.000 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.133 | 2.135 |
| Lemon sole | 0.000 | 0.002 | 0.001 | 0.992 | 0.886 | 0.078 | 0.007 | 0.000 | 0.000 | 0.000 | 0.001 | 1.968 |
| Leopardfish | 1.045 | 0.003 | 0.000 | 0.004 | 0.805 | 0.002 | 0.022 | 0.000 | 0.037 | 0.000 | 0.003 | 1.922 |
| Witch | 0.000 | 0.000 | 0.000 | 0.733 | 0.075 | 0.514 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 1.325 |
| Starry ray | 0.776 | 0.005 | 0.000 | 0.188 | 0.057 | 0.001 | 0.000 | 0.000 | 0.001 | 0.000 | 0.001 | 1.029 |
| Common dab | 0.007 | 0.002 | 0.004 | 0.574 | 0.025 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.612 |
| Halibut | 0.377 | 0.004 | 0.000 | 0.034 | 0.114 | 0.014 | 0.001 | 0.000 | 0.000 | 0.000 | 0.008 | 0.552 |
| Lumpfish | 0.000 | 0.017 | 0.001 | 0.002 | 0.002 | 0.000 | 0.037 | 0.000 | 0.000 | 0.000 | 0.333 | 0.391 |
| Megrim | 0.000 | 0.000 | 0.000 | 0.089 | 0.052 | 0.111 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.252 |
| Long rough dab | 0.009 | 0.004 | 0.000 | 0.173 | 0.031 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.220 |
| Sea-urchins | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.146 | 0.000 | 0.146 |
| European whelk | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.142 | 0.142 |
| Skate | 0.042 | 0.007 | 0.000 | 0.026 | 0.024 | 0.008 | 0.000 | 0.000 | 0.000 | 0.000 | 0.009 | 0.117 |
| Black scabbard-fish | 0.002 | 0.000 | 0.000 | 0.000 | 0.107 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.109 |
| Boston hake | 0.109 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.109 |
| Blue mussel | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.077 | 0.000 | 0.077 |
| Dogfish | 0.011 | 0.039 | 0.000 | 0.004 | 0.006 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 0.062 |
| Rat-tail | 0.000 | 0.000 | 0.000 | 0.000 | 0.058 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.059 |
| Squid | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.051 | 0.000 | 0.000 | 0.000 | 0.000 | 0.051 |
| Greenland shark | 0.000 | 0.000 | 0.000 | 0.000 | 0.043 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.043 |
| Norway pout | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.039 | 0.000 | 0.000 | 0.000 | 0.000 | 0.039 |
| onioin eye | 0.000 | 0.000 | 0.000 | 0.000 | 0.023 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.023 |
| Fuller's ray | 0.018 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.019 |
| Arctiv wolffish | 0.000 | 0.000 | 0.000 | 0.000 | 0.017 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.017 |
| sailray | 0.012 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.012 |
| Deal fish | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.011 |
| Gurnard | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.010 | 0.000 | 0.000 | 0.000 | 0.000 | 0.010 |
| Black dogfish | 0.001 | 0.000 | 0.000 | 0.000 | 0.009 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.010 |
| Total | 110.370 | 22.520 | 6.386 | 28.638 | 252.947 | 8.466 | 461.586 | 153.175 | 9.579 | 2.470 | 5.263 | 1,061.401 |

Table 7.2. Estimates of discard of cod and haddock in the Icelandic fisheries in 2008. Source: Ólafur K. Pálsson, Höskuldur Björnsson, Eypór Björnsson, Guðmundur Jóhannesson og Pórhallur Ottesen 2009. Discards in demersal Icelandic fisheries 2009. Marine Research Institute, 2009, report series no. 154 .

|  | Gear | Landings (tonnes) | Discards |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Numbers (thous.) | Weight (tonnes) | \% Weight |
| COD | Longline <br> Gillnet <br> Danish Seine <br> Bottom trawl <br> Total | $\begin{array}{r} \hline 61008 \\ 21859 \\ 10369 \\ 77172 \\ \mathbf{1 7 0 4 0 8} \end{array}$ | 509 0 28 690 1227 | 308 0 18 635 961 | $\begin{aligned} & \hline 0.51 \\ & 0.00 \\ & 0.18 \\ & 0.82 \\ & 0.56 \end{aligned}$ |
| HADDOCK | Longline <br> Danish Seine <br> Bottom trawl <br> Total | $\begin{array}{r} \hline 26573 \\ 15126 \\ 38822 \\ 808521 \\ \hline \end{array}$ | 155 36 1042 1233 | 79 9 465 553 | $\begin{aligned} & \hline 0.30 \\ & 0.06 \\ & 1.20 \\ & 0.69 \end{aligned}$ |



Figure 7.3.1. Temperature and salinity in winter west of Iceland 1971-2011. Mean 0-200m


Figure 7.3.2. Temperature and salinity off central North-Iceland 1974-2011.


Figure 7.3.3. Distribution and migrations of capelin in the Iceland/East-Greenland/Jan Mayen area before 2001. Red: Spawning grounds; Green: Adult feeding area; Blue: Distribution and feeding area of juveniles; Green arrows: Adult feeding migrations; Blue arrows: Return migrations; Red arrows: Spawning migrations; Depth contours are 200, 500 and 1000 m (Vilhjalmsson 2002)


Figure 7.3.4. Likely changes of distribution and migration routes of capelin in the Iceland/Greenland/Jan Mayen area in the last 3-4 years. Green: Feeding area; Light blue: Juvenile area; Red area: Main spawning grounds; Lighter red colour: Lesser importance of W-Iceland spawning areas; Light blue arrows: Larval drift; Dark green arrows: Feeding migrations; Dark blue arrows: Return migrations; Red arrows: Spawning migrations. Depth contours are 200, 500 and 1000 m .


Figure 7.4.1. Distribution of total catch of all species by the Icelandic fishing fleet in Icelandic EEZ and adjacent waters in 2012. The EEZs are shown as white lines.


Figure 7.4.2. Location of catches of cod, saithe, haddock, redfish, Greenland halibut and others caught with bottom trawl in 2012.


Figure 7.4.3. Location of catches of cod, saithe, haddock, redfish, Greenland halibut and others caught with long-line in 2012.


Figure 7.4.4. Location of catches of cod, saithe, haddock, redfish, Greenland halibut and others caught with gillnets in 2012.


Figure 7.4.5. Location of catches of capelin, Icelandic summer spawning herring, blue whiting and mackerel with purse seine and pelagic trawls in 2012.


Figure 7.4.6 Spatial distribution of the trawler fleet effort (in hours trawled) in 2000-2012 and as a time-series.


Figure 7.4.7. Spatial distribution of the longlinefleet effort (in number of hooks) in 2000-2012. The main targeted species for longline fishing are cod, haddock, catfish, tusk, ling and blue ling.


Figure 7.5.1. Overview of closed areas around Iceland in 2006 . The boxes are of different nature and can be closed for different time period and gear type.


Figure 7.5.2. Estimates of discard percentage by weight for cod and haddock. Source: Ólafur K. Pálsson, Höskuldur Björnsson, Eypór Björnsson, Guðmundur Jóhannesson, og Pórhallur Ottesen 2009. Discards in demersal Icelandic fisheries 2009. Marine Research Institute, report series Nr. 154. 2010 figures are preliminary .


Figure 7.6.1. Cumulative plot for bottom trawl in 2008. An example describes this probably best. Looking at the figure above it can be seen from the dashed lines that $30 \%$ of the catch of haddock comes from hauls where haddock is less than $60 \%$ of the total catch while only $4 \%$ of the catch of greenland halibut comes from hauls where it is less than $50 \%$ of the total catch. $75 \%$ of the plaice is on the other hand caught in hauls where plaice is minority of the catches. The figures also shows that $70 \%$ of the catch of greenland halibut comes from hauls where nothing else is caught but only $10 \%$ of the haddock. Of the species shown in the figure plaice is the one with largest proportion as bycatch while greenland halibut is the one with largest proportion caught in mixed fisheries.

## Summary

The 2015 reference biomass $\left(B_{4+}\right)$ is estimated as 255 kt , around the average in the assessment period (1980 to the present). Spawning biomass is estimated as 139 kt , above the average in the assessment period and well above $B_{\text {trigger }}=65 \mathrm{kt}$ and $B_{\lim }=61 \mathrm{kt}$.

Harvest rate has been around the HCR target of $20 \%$ since 2011, with fishing mortality rate between 0.19-0.25. Year classes 2008 and 2009 are above average, but recruitment has declined below average since then.

Weights of ages 3-6 have been low in recent years, but older ages are close to average weight. Maturity-at-ages 4-9 has decreased in recent years and is currently around average.

The assessment model is a separable statistical catch-at-age model implemented in AD Model Builder. Selectivity is age-specific and varies between three periods: 1980-1996, 1997-2003, and 2004 onwards.

The default separable model (ADSEP) estimates a slightly larger stock size than alternative diagnostic models (ADAPT, TSA, SAM). The estimates of this year's B4+ range from 209 (TSA) to 255 kt (ADSEP).
In 2013, the Icelandic government adopted a harvest control rule for managing the Icelandic saithe fishery, evaluated by ICES (2013). It is similar to the $20 \%$ rule used for the Icelandic cod fishery. When the population is above $B_{\text {trigger, }}$ the TAC set in year $t$ equals the average of $0.2 B_{4+}$ in year $t$ and last year's TAC.
According to the adopted harvest control rule, the TAC will be 55 kt in the next fishing year.

### 8.1 Stock description and management units

Description of the stock and management units is provided in the stock annex.

### 8.2 Fisheries-dependent data

### 8.2.1 Landings, advice and TAC

Landings of saithe in Icelandic waters in 2014 are estimated to have been 46500 t (Table 8.1 and Figure 8.1). Of the landings, 38600 t were caught by trawl, 2400 t by gillnets, and the rest caught by other fishing gear. The domestic as well as ICES advice for the fishing year 2014/2015 was based on the $20 \%$ harvest control rule and was 58 kt . The TAC issued was also 58 kt . The trajectory of the landings in the current fishing year and calendar year is shown in Figure 8.2.

Most of the catch is caught in bottom trawl ( $80 \%$ in 2010-2014), with gillnet and jiggers taking the majority of the rest. The share taken by the gillnet fleet was larger in the past, $26 \%$ in 19821996 compared to $9 \%$ in 19972014 (Figure 8.1).

### 8.2.2 Landings by age

Catch in numbers by age based on landings are listed in Table 8.2. Discarding is not considered to be a problem in the Icelandic saithe fisheries, with an estimated discard proportion of $0.1 \%$ (annual reports by Palsson et al. 2003 and later). Comparison of sea and harbour samples indicate that discards have been small in most years since 2000.

The sea samples constitute about $6070 \%$ of the length samples used in the calculation of the catch in number. Since the amount of discards is likely to be small, not taking discards into account in the total catches and catch in numbers is not considered to have major effect on the stock assessment.
The sampling program was slightly revised in 2013 and 2014, but the approach used for calculating catch in numbers has not changed. In 2013, the sampling frequency was reduced for bottom trawl, while the sampling frequency was increased for gillnets, jiggers, and demersal seine in 2014. Also in 2014, the number of otoliths from each sample was halved from 5025 for all fishing gears. These revisions in the sampling program were based on the analysis of Thordarson (2012). The age and length sampling in 2014 is indicated in the following table:

| Fleet | LANDINGS <br> (T) | No. of OTOLITH SAMPLES | No. OF OTOLITHS READ | No. of <br> LENGTH <br> SAMPLES | No. of Length MEASUREMENTS |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Gillnets | 2355 | 9 | 250 | 10 | 1036 |
| Jiggers | 2115 | 14 | 370 | 15 | 1601 |
| Demersal seine | 1005 | 4 | 150 | 4 | 471 |
| Bottom trawl | 38634 | 52 | 1625 | 224 | 32251 |
| Other gear | 1624 | - | - | 189 | 2354 |
| Foreign landings | 750 | - | - | - | - |
| Total | 46483 | 79 | 2395 | 442 | 37713 |

Two age-length keys are used to calculate catch-at-age, one key for the gillnet catch and another key for other gears combined. The same length-weight relationship ( $W=0.02498$ * $L^{\wedge} 2.75674$ ) is applied to length distributions from both fleets.

### 8.2.3 Mean weight and maturity-at-age

Weights of ages 3-6 have been low in recent years, but older ages are close to average weight (Table 8.3 and Figure 8.3). The long-term trend since 1980 has been a gradual decline in the weight of all ages. Weight at age in the landings is also used as weight at age in the stock. Weights for the current calendar year are predicted by applying a linear model using survey weights and the weight of a year class in the previous year as predictors (Magnusson 2012).
Maturity-at-ages 49 has decreased in recent years and is currently around average (Table 8.4 and Figure 8.4). A model using maturity-at-age from the Icelandic groundfish spring survey is used to derive smoothed trends in maturity by age and year (see stock annex).

### 8.2.4 Logbook data

Commercial cpue indices are not used for tuning in this assessment. Although these indices have been explored for inclusion in the past, they were not considered for inclusion in the benchmark (ICES 2010), as the trends in cpue are considered unreliable as an indicator of changes in abundance.

### 8.3 Scientific surveys

In the benchmark, spring survey data were considered superior to the autumn survey for calibrating the assessment. Saithe is among the most difficult demersal fish to get reliable information on from bottom trawl surveys. In the spring survey, which has

500600 stations, a large proportion of the saithe is caught in relatively few hauls and there seems to be considerable interannual variability of the number of these hauls.

The survey biomass indices fluctuated greatly in 19851995, but were consistently low in 1995-2001, high in the period around 2005, declining to a relatively low level in 20072011. The 2012 and 2013 survey biomass indices were relatively high (Table 8.5 and Figure 8.5).

Internal consistency in the surveys measured by the correlation of the indices for the same year class in 2 adjacent surveys is poor, with $R^{2}$ close to 0.3 for the best-defined age groups, and much lower for some other.
Young saithe tend to live very close to shore, so it is not surprising that survey indices for ages 1 and 2 are poor measures of recruitment, and the number of young saithe caught in the survey is very low.

### 8.4 Assessment method

In accordance with the recommendation from the benchmark (ICES 2010), a separable forward-projecting statistical catch-age model, developed in AD Model Builder, is used to fit commercial catch-at-age (ages 314 from 1980 onwards) and survey catch-atage (ages 210 from 1985 onwards). The selectivity pattern is constant within each period (Figure 8.6). Natural mortality is set at 0.2 for all ages.

The commercial catch-at-age residuals (Table 8.6 and Figure 8.7) are relatively small in recent years, owing to the model flexibility provided by the two recent selectivity periods 19972003 and 2004 onwards. The survey catch-at-age residuals (Table 8.7 and Figure 8.7) have year blocks with all residuals being only negative or only positive in some years. The survey residuals are modelled as multivariate normal distribution with the correlation estimated (one coefficient).

### 8.5 Reference points and HCR

In April 2013, the Icelandic government adopted a management plan for managing the Icelandic saithe fishery (Ministry of Industries and Innovation 2013). ICES evaluated this management plan and concluded that it was in accordance with the precautionary approach and the ICES MSY framework. In the harvest control rule (HCR) evaluation (ICES 2013) $B$ lim was defined as 61 kt , based on $B_{\text {loss }}$ as estimated in 2010, and $B_{\text {trigger }}$ was defined as 65 kt , based on an estimated hockey-stick recruitment function.
The TAC set in year $t$ is for the upcoming fishing year, from 1 September in year $t$, to 31 August in year $t+1$. The $20 \%$ HCR consists of two equations, as follows.
When $S S B \geq B_{\text {trigger, }}$ the TAC set in year $t$ equals the average of 0.20 times the current biomass and last year's TAC:

$$
\begin{equation*}
T A C_{t}=0.5 \times 0.20 B_{t, 4+}+0.5 T A C_{t-1} \tag{Eq.1}
\end{equation*}
$$

When SSB is below $B_{\text {trigger, }}$ the harvest rate is reduced below 0.20 :

$$
\begin{equation*}
\left.T A C_{\mathrm{t}}=S S B_{t} / B_{\text {trigger }}\left[\left(1-0.5 S S B_{t} / B_{\text {trigger }}\right) 0.20 B_{t, 4+}\right)+0.5 T A C_{t-1}\right] \tag{Eq.2}
\end{equation*}
$$

Equation 1 is a plain average of two numbers. Equation 2 is continuous over $S S B_{t} / B_{\text {trig }}$ ger, so the rule does not lead to very different TAC when $S S B_{t}$ is slightly below or above $B_{\text {trigger }}$ (Magnusson 2013).

### 8.6 State of the stock

The results of the principal stock quantities (Table 8.8 and Figure 8.8) show that the reference biomass has historically ranged from 410130 kt (in 1988 and 1999), but this range has been narrower since 2003, between 220320 kt . The current stock size of 255 kt is around the average in the assessment period ( 1980 to the present). Spawning biomass is estimated as 139 kt , above the average in the assessment period and well above $B_{\text {trigger }}$ and $B$ lim.
The harvest rate peaked around $30 \%$ in the mid- 1990 s, but has fluctuated around the HCR target of $20 \%$ since 2011, with fishing mortality rate between 0.19 and 0.25 . SSB has been stable at a relatively high level during the last ten years, having declined to its historical minimum in the mid-1990s.

Year classes 2008 and 2009 are above average, but recruitment has declined below average since then. The details of the fishing mortality and stock in numbers are presented in Tables 8.9 and 8.10.

### 8.7 Short-term forecast

The input for the short-term forecast is shown in Table 8.11. Future weights, maturity, and selectivity are assumed to be the same as in the assessment year, as described in the stock annex. Recruitment predictions are based on the segmented stock-recruitment function estimated in the assessment model.

The landings for the ongoing calendar year are predicted based on the $20 \% \mathrm{HCR}$, with the calendar year landings consisting of $2 / 3$ of the ongoing fishing year's TAC and $1 / 3$ of the next fishing year's TAC.
Following the HCR, the predicted landings in 2016 are 54 kt and the resulting SSB in 2017 is predicted to be 130 kt .

### 8.8 Uncertainties in assessment and forecast

The assessment of Icelandic saithe is relatively uncertain due to fluctuations in the survey data, as well as irregular changes in the fleet selectivity. The internal consistency in the spring bottom-trawl survey is very low for saithe. This is not surprising, considering the nature of the species that is partly pelagic, schooling, and relatively widely migrating. There are also indications of time-varying selectivity, so changes in the commercial catch-at-age may not reflect changes in the age distribution of the population. The retrospective pattern (Figure 8.9) reveals some of the assessment uncertainty. The harvest control rule evaluation incorporated uncertainties about assessment estimates, among other sources of uncertainty (ICES 2013).

The results from the default separable assessment model (ADSEP) are compared to alternative diagnostic model runs, involving ADAPT, TSA, and SAM, in order to explore the overall uncertainty in the assessment. The comparison involved four models which differ mainly in the way the commercial catch-at-age variability and F-matrix is modelled:

|  | MODEL | FAMILY | CA variability | F matrix |
| :--- | :--- | :--- | :--- | :--- |
| 1 | ADSEP <br> (default) | separable | observation <br> error | multiplicative <br> in 3 periods |
| 2 | ADAPT | vpa | process error | no constraints |
| 3 | TSA | state-space <br> (kalman filter) | observation \& process <br> error | orthogonal <br> polynomials |
| 4 | SAM | state-space <br> (random effects) | observation \& process <br> error | correlated <br> random walk |

The results from the model comparison (Figure 8.10) show that the default model estimates a slightly larger stock size than the other models, which has also been the case for saithe assessments in recent years. The estimates of this year's $B_{4+}$ range from 209 (TSA) to 255 kt (ADSEP).

### 8.9 Comparison with previous assessment and forecast

Compared to last year's assessment the estimated reference biomass B4+ in 2014 has decreased from 296265 kt , SSB 2014 has decreased from 150132 kt , and the harvest rate u2013 has increased from $19 \%$ to $22 \%$ (fishing mortality 0.220 .25 ). Stock numbers-at-age 5 have increased slightly, while stock numbers-at-ages 6 and 7 have decreased as shown below.

|  |  | NWWG 2014 | NWWG 2015 |
| :--- | :--- | :--- | :--- |
| B4+(2014) | 296 | 265 |  |
| SSB(2014) | 150 | 132 |  |
| $\mathrm{u}(2013)$ | $19 \%$ | $22 \%$ |  |
| F4-9(2013) | 0.22 | 0.25 |  |
| N5(2014) | 24 | 26 |  |
| N6(2014) | 21 | 17 |  |
| N7(2014) | 11 | 9 |  |

### 8.10 Ecosystem considerations

Changes in the distribution of large pelagic stocks (blue whiting, mackerel, Norwegian spring-spawning herring, Icelandic summer-spawning herring) may affect the propensity of saithe to migrate off shelf and between management units. Saithe is a migrating species and makes both vertical and long-distance feeding and spawning migrations (Armannsson et al. 2007, Armannsson and Jonsson 2012, Homrum et al. 2013). The evidence from tagging experiments (ICES 2008) show some migrations along the FaroeIceland Ridge, as well as onto the East Greenland shelf.

### 8.11 Changes in fishing technology and fishing patterns

According to the stock assessment model fit to the commercial catch-at-age data, the fleet is targeting younger fish since around 2004, compared to earlier years. This can be partly explained by reduced use of gillnets in the saithe fishery.

### 8.12 References

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Table 8.1. Saithe in division Va. Nominal catch ( $t$ ) by countries, as officially reported to ICES.

|  | Belgium | Faroes | France | Germany | ICELAND | Norway | UK (E/W/NI) | UK (SCOT) | UK | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 980 | 4930 |  |  | 52436 | 1 |  |  |  | 58347 |
| 1981 | 532 | 3545 |  |  | 54921 | 3 |  |  |  | 59001 |
| 1982 | 201 | 3582 | 23 |  | 65124 | 1 |  |  |  | 68931 |
| 1983 | 224 | 2138 |  |  | 55904 |  |  |  |  | 58266 |
| 1984 | 269 | 2044 |  |  | 60406 |  |  |  |  | 62719 |
| 1985 | 158 | 1778 |  |  | 55135 | 1 | 29 |  |  | 57101 |
| 1986 | 218 | 2291 |  |  | 63867 |  |  |  |  | 66376 |
| 1987 | 217 | 2139 |  |  | 78175 |  |  |  |  | 80531 |
| 1988 | 268 | 2596 |  |  | 74383 |  |  |  |  | 77247 |
| 1989 | 369 | 2246 |  |  | 79796 |  |  |  |  | 82411 |
| 1990 | 190 | 2905 |  |  | 95032 |  |  |  |  | 98127 |
| 1991 | 236 | 2690 |  |  | 99811 |  |  |  |  | 102737 |
| 1992 | 195 | 1570 |  |  | 77832 |  |  |  |  | 79597 |
| 1993 | 104 | 1562 |  |  | 69982 |  |  |  |  | 71648 |
| 1994 | 30 | 975 |  | 1 | 63333 |  |  |  |  | 64339 |
| 1995 |  | 1161 |  | 1 | 47466 | 1 |  |  |  | 48629 |
| 1996 |  | 803 |  | 1 | 39297 |  |  |  |  | 40101 |
| 1997 |  | 716 |  |  | 36548 |  |  |  |  | 37264 |
| 1998 |  | 997 |  | 3 | 30531 |  |  |  |  | 31531 |
| 1999 |  | 700 |  | 2 | 30583 | 6 | 1 | 1 |  | 31293 |
| 2000 |  | 228 |  | 1 | 32914 | 1 | 2 |  |  | 33146 |
| 2001 |  | 128 |  | 14 | 31854 | 44 | 23 |  |  | 32063 |
| 2002 |  | 366 |  | 6 | 41687 | 3 | 7 | 2 |  | 42071 |
| 2003 |  | 143 |  | 56 | 51857 | 164 |  |  | 35 | 52255 |
| 2004 |  | 214 |  | 157 | 62614 | 1 | 105 |  |  | 63091 |
| 2005 |  | 322 |  | 224 | 67283 | 2 |  |  | 312 | 68143 |
| 2006 |  | 415 |  | 33 | 75197 | 2 |  |  | 16 | 75663 |
| 2007 |  | 392 |  |  | 64008 | 3 |  |  | 30 | 64433 |
| 2008 |  | 196 |  |  | 69992 | 2 |  |  |  | 70190 |
| 2009 |  | 269 |  |  | 61391 | 3 |  |  |  | 61663 |
| 2010 |  | 499 |  |  | 53772 | 1 |  |  |  | 54272 |
| 2011 |  | 735 |  |  | 50386 | 2 |  |  |  | 51123 |
| 2012 |  | 940 |  |  | 50843 |  |  |  |  | 51783 |
| 2013 |  | 925 |  |  | 57077 |  |  |  |  | 58002 |
| 2014 |  | 746 |  |  | 45733 | 4 |  |  |  | 46483 |

Table 8.2. Saithe in division Va. Commercial catch-at-age (millions).

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.275 | 2.540 | 5.214 | 2.596 | 2.169 | 1.341 | 0.387 | 0.262 | 0.155 | 0.112 | 0.064 | 0.033 |
| 1981 | 0.203 | 1.325 | 3.503 | 5.404 | 1.457 | 1.415 | 0.578 | 0.242 | 0.061 | 0.154 | 0.135 | 0.128 |
| 1982 | 0.508 | 1.092 | 2.804 | 4.845 | 4.293 | 1.215 | 0.975 | 0.306 | 0.059 | 0.035 | 0.048 | 0.046 |
| 1983 | 0.107 | 1.750 | 1.065 | 2.455 | 4.454 | 2.311 | 0.501 | 0.251 | 0.038 | 0.012 | 0.002 | 0.004 |
| 1984 | 0.053 | 0.657 | 0.800 | 1.825 | 2.184 | 3.610 | 0.844 | 0.376 | 0.291 | 0.135 | 0.185 | 0.226 |
| 1985 | 0.376 | 4.014 | 3.366 | 1.958 | 1.536 | 1.172 | 0.747 | 0.479 | 0.074 | 0.023 | 0.072 | 0.071 |
| 1986 | 3.108 | 1.400 | 4.170 | 2.665 | 1.550 | 1.116 | 0.628 | 1.549 | 0.216 | 0.051 | 0.030 | 0.014 |
| 1987 | 0.956 | 5.135 | 4.428 | 5.409 | 2.915 | 1.348 | 0.661 | 0.496 | 0.498 | 0.058 | 0.027 | 0.048 |
| $1988$ | 1.318 | 5.067 | 6.619 | 3.678 | 2.859 | 1.775 | 0.845 | 0.226 | 0.270 | 0.107 | 0.024 | 0.001 |
| 1989 | 0.315 | 4.313 | 8.471 | 7.309 | 1.794 | 1.928 | 0.848 | 0.270 | 0.191 | 0.135 | 0.076 | 0.010 |
| 1990 | 0.143 | 1.692 | 5.471 | 10.112 | 6.174 | 1.816 | 1.087 | 0.380 | 0.151 | 0.055 | 0.076 | 0.037 |
| 1991 | 0.198 | 0.874 | 3.613 | 6.844 | 10.772 | 3.223 | 0.858 | 0.838 | 0.228 | 0.040 | 0.006 | 0.005 |
| 1992 | 0.242 | 2.928 | 3.844 | 4.355 | 3.884 | 4.046 | 1.290 | 0.350 | 0.196 | 0.056 | 0.054 | 0.015 |
| 1993 | 0.657 | 1.083 | 2.841 | 2.252 | 2.247 | 2.314 | 3.671 | 0.830 | 0.223 | 0.188 | 0.081 | 0.012 |
| 1994 | 0.702 | 2.955 | 1.770 | 2.603 | 1.377 | 1.243 | 1.263 | 2.009 | 0.454 | 0.158 | 0.188 | 0.082 |
| 1995 | 1.573 | 1.853 | 2.661 | 1.807 | 2.370 | 0.905 | 0.574 | 0.482 | 0.521 | 0.106 | 0.035 | 0.013 |
| 1996 | 1.102 | 2.608 | 1.868 | 1.649 | 0.835 | 1.233 | 0.385 | 0.267 | 0.210 | 0.232 | 0.141 | 0.074 |
| 1997 | 0.603 | 2.960 | 2.766 | 1.651 | 1.178 | 0.599 | 0.454 | 0.125 | 0.095 | 0.114 | 0.077 | 0.043 |
| 1998 | 0.183 | 1.289 | 1.767 | 1.545 | 1.114 | 0.658 | 0.351 | 0.265 | 0.120 | 0.081 | 0.085 | 0.085 |
| 1999 | 0.989 | 0.732 | 1.564 | 2.176 | 1.934 | 0.669 | 0.324 | 0.140 | 0.072 | 0.025 | 0.028 | 0.022 |
| 2000 | 0.850 | 2.383 | 0.896 | 1.511 | 1.612 | 1.806 | 0.335 | 0.173 | 0.057 | 0.033 | 0.017 | 0.007 |
| 2001 | 1.223 | 2.619 | 2.184 | 0.591 | 0.977 | 0.943 | 0.819 | 0.186 | 0.094 | 0.028 | 0.028 | 0.013 |
| 2002 | 1.187 | 4.190 | 3.147 | 2.970 | 0.519 | 0.820 | 0.570 | 0.309 | 0.101 | 0.027 | 0.015 | 0.011 |
| 2003 | 2.284 | 4.363 | 4.031 | 2.472 | 1.942 | 0.285 | 0.438 | 0.289 | 0.196 | 0.028 | 0.029 | 0.015 |
| 2004 | 0.952 | 7.841 | 7.195 | 5.363 | 1.563 | 1.057 | 0.211 | 0.224 | 0.157 | 0.074 | 0.039 | 0.011 |
| 2005 | 2.607 | 3.089 | 7.333 | 6.876 | 3.592 | 0.978 | 0.642 | 0.119 | 0.149 | 0.089 | 0.046 | 0.012 |
| 2006 | 1.380 | 10.051 | 2.616 | 5.840 | 4.514 | 1.989 | 0.667 | 0.485 | 0.118 | 0.112 | 0.086 | 0.031 |
| 2007 | 1.244 | 6.552 | 8.751 | 2.124 | 2.935 | 1.817 | 0.964 | 0.395 | 0.190 | 0.043 | 0.036 | 0.020 |
| 2008 | 1.432 | 3.602 | 5.874 | 6.706 | 1.155 | 1.894 | 1.248 | 0.803 | 0.262 | 0.176 | 0.087 | 0.044 |
| 2009 | 2.820 | 5.166 | 2.084 | 2.734 | 2.883 | 0.777 | 1.101 | 0.847 | 0.555 | 0.203 | 0.134 | 0.036 |
| 2010 | 2.146 | 6.284 | 3.058 | 0.997 | 1.644 | 1.571 | 0.514 | 0.656 | 0.522 | 0.231 | 0.114 | 0.064 |
| 2011 | 2.004 | 4.850 | 4.006 | 1.502 | 0.677 | 1.065 | 1.145 | 0.323 | 0.433 | 0.244 | 0.150 | 0.075 |
| 2012 | 1.183 | 4.816 | 3.514 | 2.417 | 0.903 | 0.432 | 0.883 | 1.015 | 0.354 | 0.277 | 0.173 | 0.099 |
| 2013 | 1.163 | 5.538 | 6.366 | 2.963 | 1.610 | 0.664 | 0.375 | 0.537 | 0.460 | 0.124 | 0.118 | 0.078 |
| 2014 | 0.668 | 3.499 | 4.867 | 2.805 | 1.276 | 0.725 | 0.347 | 0.241 | 0.312 | 0.199 | 0.128 | 0.074 |

Table 8.3. Saithe in division Va. Mean weight at age (g) in the catches and in the spawning stock, with predictions in grey.

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 1428 | 1983 | 2667 | 3689 | 5409 | 6321 | 7213 | 8565 | 9147 | 9617 | 10066 | 11041 |
| 1981 | 1585 | 2037 | 2696 | 3525 | 4541 | 6247 | 6991 | 8202 | 9537 | 9089 | 9351 | 10225 |
| 1982 | 1547 | 2194 | 3015 | 3183 | 5114 | 6202 | 7256 | 7922 | 8924 | 10134 | 9447 | 10535 |
| 1983 | 1530 | 2221 | 3171 | 4270 | 4107 | 5984 | 7565 | 8673 | 8801 | 9039 | 11138 | 9818 |
| 1984 | 1653 | 2432 | 3330 | 4681 | 5466 | 4973 | 7407 | 8179 | 8770 | 8831 | 11010 | 11127 |
| 1985 | 1609 | 2172 | 3169 | 3922 | 4697 | 6411 | 6492 | 8346 | 9401 | 10335 | 11027 | 10644 |
| 1986 | 1450 | 2190 | 2959 | 4402 | 5488 | 6406 | 7570 | 6487 | 9616 | 10462 | 11747 | 11902 |
| 1987 | 1516 | 1715 | 2670 | 3839 | 5081 | 6185 | 7330 | 8025 | 7974 | 9615 | 12246 | 11656 |
| 1988 | 1261 | 2017 | 2513 | 3476 | 4719 | 5932 | 7523 | 8439 | 8748 | 9559 | 10824 | 14099 |
| 1989 | 1403 | 2021 | 2194 | 3047 | 4505 | 5889 | 7172 | 8852 | 10170 | 10392 | 12522 | 11923 |
| 1990 | 1647 | 1983 | 2566 | 3021 | 4077 | 5744 | 7038 | 7564 | 8854 | 10645 | 11674 | 11431 |
| 1991 | 1224 | 1939 | 2432 | 3160 | 3634 | 4967 | 6629 | 7704 | 9061 | 9117 | 10922 | 11342 |
| 1992 | 1269 | 1909 | 2578 | 3288 | 4150 | 4865 | 6168 | 7926 | 8349 | 9029 | 11574 | 9466 |
| 1993 | 1381 | 2143 | 2742 | 3636 | 4398 | 5421 | 5319 | 7006 | 8070 | 10048 | 9106 | 11591 |
| 1994 | 1444 | 1836 | 2649 | 3512 | 4906 | 5539 | 6818 | 6374 | 8341 | 9770 | 10528 | 11257 |
| 1995 | 1370 | 1977 | 2769 | 3722 | 4621 | 5854 | 6416 | 7356 | 6815 | 8312 | 9119 | 11910 |
| 1996 | 1229 | 1755 | 2670 | 3802 | 4902 | 5681 | 7182 | 7734 | 9256 | 8322 | 10501 | 11894 |
| 1997 | 1325 | 1936 | 2409 | 3906 | 5032 | 6171 | 7202 | 7883 | 8856 | 9649 | 9621 | 10877 |
| 1998 | 1347 | 1972 | 2943 | 3419 | 4850 | 5962 | 6933 | 7781 | 8695 | 9564 | 10164 | 10379 |
| 1999 | 1279 | 2106 | 2752 | 3497 | 3831 | 5819 | 7072 | 8078 | 8865 | 10550 | 10823 | 11300 |
| 2000 | 1367 | 1929 | 2751 | 3274 | 4171 | 4447 | 6790 | 8216 | 9369 | 9817 | 10932 | 12204 |
| 2001 | 1280 | 1882 | 2599 | 3697 | 4420 | 5538 | 5639 | 7985 | 9059 | 9942 | 10632 | 10988 |
| 2002 | 1308 | 1946 | 2569 | 3266 | 4872 | 5365 | 6830 | 7067 | 9240 | 9659 | 10088 | 11632 |
| 2003 | 1310 | 1908 | 2545 | 3336 | 4069 | 5792 | 7156 | 8131 | 8051 | 10186 | 10948 | 11780 |
| 2004 | 1467 | 1847 | 2181 | 2918 | 4017 | 5135 | 7125 | 7732 | 8420 | 8927 | 10420 | 10622 |
| 2005 | 1287 | 1888 | 2307 | 2619 | 3516 | 5080 | 6060 | 8052 | 8292 | 8342 | 8567 | 10256 |
| 2006 | 1164 | 1722 | 2369 | 2808 | 3235 | 4361 | 6007 | 7166 | 8459 | 9324 | 9902 | 9636 |
| 2007 | 1140 | 1578 | 2122 | 2719 | 3495 | 4114 | 5402 | 6995 | 7792 | 9331 | 9970 | 10738 |
| 2008 | 1306 | 1805 | 2295 | 2749 | 3515 | 4530 | 5132 | 6394 | 7694 | 9170 | 9594 | 11258 |
| 2009 | 1412 | 1862 | 2561 | 3023 | 3676 | 4596 | 5651 | 6074 | 7356 | 8608 | 9812 | 10639 |
| 2010 | 1287 | 1787 | 2579 | 3469 | 4135 | 4850 | 5558 | 6289 | 6750 | 7997 | 9429 | 10481 |
| 2011 | 1175 | 1801 | 2526 | 3680 | 4613 | 5367 | 5685 | 6466 | 6851 | 7039 | 8268 | 8958 |
| 2012 | 1160 | 1668 | 2369 | 3347 | 4430 | 5486 | 6161 | 6448 | 7220 | 8054 | 8147 | 8901 |
| 2013 | 1056 | 1675 | 2219 | 3244 | 4529 | 5628 | 6397 | 7055 | 7378 | 7955 | 8400 | 8870 |
| 2014 | 1211 | 1575 | 2229 | 2983 | 4378 | 5598 | 6773 | 8023 | 7875 | 8646 | 9179 | 9749 |
| 2015 | 1142 | 1726 | 2217 | 3071 | 4030 | 5532 | 6846 | 7175 | 7491 | 8218 | 8575 | 9173 |
| 2016 | 1142 | 1726 | 2217 | 3071 | 4030 | 5532 | 6846 | 7175 | 7491 | 8218 | 8575 | 9173 |
| 2017 | 1142 | 1726 | 2217 | 3071 | 4030 | 5532 | 6846 | 7175 | 7491 | 8218 | 8575 | 9173 |

Table 8.4. Saithe in division Va. Maturity-at-age used for calculating the SSB.

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0 | 0.089 | 0.197 | 0.380 | 0.604 | 0.792 | 0.905 | 1 | 1 | 1 | 1 | 1 |
| 1986 | 0 | 0.080 | 0.178 | 0.351 | 0.575 | 0.772 | 0.894 | 1 | 1 | 1 | 1 | 1 |
| 1987 | 0 | 0.072 | 0.162 | 0.325 | 0.547 | 0.751 | 0.883 | 1 | 1 | 1 | 1 | 1 |
| 1988 | 0 | 0.065 | 0.148 | 0.303 | 0.521 | 0.731 | 0.871 | 1 | 1 | 1 | 1 | 1 |
| 1989 | 0 | 0.060 | 0.138 | 0.285 | 0.499 | 0.714 | 0.862 | 1 | 1 | 1 | 1 | 1 |
| 1990 | 0 | 0.057 | 0.131 | 0.273 | 0.484 | 0.701 | 0.854 | 1 | 1 | 1 | 1 | 1 |
| 1991 | 0 | 0.055 | 0.127 | 0.266 | 0.475 | 0.694 | 0.850 | 1 | 1 | 1 | 1 | 1 |
| 1992 | 0 | 0.055 | 0.127 | 0.266 | 0.476 | 0.694 | 0.850 | 1 | 1 | 1 | 1 | 1 |
| 1993 | 0 | 0.057 | 0.131 | 0.274 | 0.485 | 0.702 | 0.855 | 1 | 1 | 1 | 1 | 1 |
| 1994 | 0 | 0.062 | 0.141 | 0.290 | 0.505 | 0.718 | 0.864 | 1 | 1 | 1 | 1 | 1 |
| 1995 | 0 | 0.069 | 0.157 | 0.317 | 0.537 | 0.743 | 0.879 | 1 | 1 | 1 | 1 | 1 |
| 1996 | 0 | 0.081 | 0.181 | 0.355 | 0.579 | 0.775 | 0.896 | 1 | 1 | 1 | 1 | 1 |
| 1997 | 0 | 0.097 | 0.212 | 0.402 | 0.627 | 0.807 | 0.913 | 1 | 1 | 1 | 1 | 1 |
| 1998 | 0 | 0.117 | 0.248 | 0.451 | 0.673 | 0.837 | 0.928 | 1 | 1 | 1 | 1 | 1 |
| 1999 | 0 | 0.137 | 0.284 | 0.497 | 0.712 | 0.860 | 0.939 | 1 | 1 | 1 | 1 | 1 |
| 2000 | 0 | 0.154 | 0.313 | 0.532 | 0.740 | 0.877 | 0.947 | 1 | 1 | 1 | 1 | 1 |
| 2001 | 0 | 0.165 | 0.331 | 0.552 | 0.755 | 0.885 | 0.951 | 1 | 1 | 1 | 1 | 1 |
| 2002 | 0 | 0.169 | 0.337 | 0.560 | 0.760 | 0.888 | 0.952 | 1 | 1 | 1 | 1 | 1 |
| 2003 | 0 | 0.168 | 0.335 | 0.557 | 0.759 | 0.887 | 0.952 | 1 | 1 | 1 | 1 | 1 |
| 2004 | 0 | 0.163 | 0.328 | 0.549 | 0.753 | 0.884 | 0.950 | 1 | 1 | 1 | 1 | 1 |
| 2005 | 0 | 0.157 | 0.318 | 0.538 | 0.744 | 0.879 | 0.948 | 1 | 1 | 1 | 1 | 1 |
| 2006 | 0 | 0.152 | 0.309 | 0.527 | 0.736 | 0.874 | 0.946 | 1 | 1 | 1 | 1 | 1 |
| 2007 | 0 | 0.146 | 0.300 | 0.517 | 0.728 | 0.870 | 0.943 | 1 | 1 | 1 | 1 | 1 |
| 2008 | 0 | 0.141 | 0.291 | 0.506 | 0.719 | 0.865 | 0.941 | 1 | 1 | 1 | 1 | 1 |
| 2009 | 0 | 0.136 | 0.282 | 0.495 | 0.710 | 0.859 | 0.939 | 1 | 1 | 1 | 1 | 1 |
| 2010 | 0 | 0.130 | 0.272 | 0.483 | 0.700 | 0.853 | 0.936 | 1 | 1 | 1 | 1 | 1 |
| 2011 | 0 | 0.124 | 0.261 | 0.469 | 0.688 | 0.847 | 0.932 | 1 | 1 | 1 | 1 | 1 |
| 2012 | 0 | 0.118 | 0.250 | 0.455 | 0.676 | 0.839 | 0.929 | 1 | 1 | 1 | 1 | 1 |
| 2013 | 0 | 0.112 | 0.239 | 0.440 | 0.662 | 0.830 | 0.924 | 1 | 1 | 1 | 1 | 1 |
| 2014 | 0 | 0.106 | 0.228 | 0.424 | 0.648 | 0.821 | 0.920 | 1 | 1 | 1 | 1 | 1 |
| 2015 | 0 | 0.100 | 0.217 | 0.409 | 0.633 | 0.812 | 0.915 | 1 | 1 | 1 | 1 | 1 |
| 2016 | 0 | 0.100 | 0.217 | 0.409 | 0.633 | 0.812 | 0.915 | 1 | 1 | 1 | 1 | 1 |
| 2017 | 0 | 0.100 | 0.217 | 0.409 | 0.633 | 0.812 | 0.915 | 1 | 1 | 1 | 1 | 1 |

Table 8.5. Saithe in division Va. Survey catch-at-age.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0.05 | 0.61 | 0.58 | 2.99 | 5.11 | 1.74 | 1.06 | 0.50 | 1.37 | 0.16 | 0.08 | 0.08 | 0.07 | 0.07 |
| 1986 | 0.02 | 2.33 | 2.40 | 2.06 | 2.09 | 1.42 | 0.62 | 0.28 | 0.19 | 0.32 | 0.09 | 0.07 | 0.03 | 0.00 |
| 1987 | 0.10 | 0.39 | 11.52 | 12.93 | 6.42 | 3.95 | 3.07 | 0.79 | 0.36 | 0.26 | 0.33 | 0.05 | 0.01 | 0.03 |
| 1988 | 0.69 | 0.31 | 0.49 | 2.72 | 2.81 | 1.71 | 0.95 | 0.40 | 0.07 | 0.08 | 0.10 | 0.05 | 0.01 | 0.00 |
| 1989 | 0.20 | 1.43 | 3.96 | 5.05 | 6.57 | 2.49 | 1.77 | 0.91 | 0.40 | 0.00 | 0.02 | 0.00 | 0.03 | 0.00 |
| 1990 | 0.01 | 0.35 | 1.69 | 4.86 | 6.37 | 12.33 | 3.30 | 1.21 | 0.64 | 0.12 | 0.06 | 0.02 | 0.01 | 0.03 |
| 1991 | 0.01 | 0.22 | 1.40 | 1.72 | 2.22 | 1.13 | 2.50 | 0.30 | 0.02 | 0.03 | 0.00 | 0.01 | 0.00 | 0.01 |
| 1992 | 0.01 | 0.15 | 0.91 | 5.73 | 5.52 | 2.79 | 2.68 | 1.91 | 0.28 | 0.06 | 0.06 | 0.02 | 0.00 | 0.00 |
| 1993 | $0.00$ | 1.27 | 11.04 | 2.00 | 6.80 | 2.41 | 2.25 | 1.02 | 4.02 | 0.64 | 0.05 | 0.00 | 0.02 | 0.00 |
| 1994 | 0.04 | 0.82 | 0.73 | 1.89 | 1.74 | 1.95 | 0.53 | 0.84 | 1.00 | 3.62 | 0.41 | 0.18 | 0.00 | 0.04 |
| 1995 | 0.06 | 0.48 | 1.98 | 1.12 | 0.51 | 0.28 | 0.34 | 0.10 | 0.15 | 0.15 | 0.33 | 0.02 | 0.00 | 0.00 |
| 1996 | 0.03 | 0.13 | 0.51 | 3.76 | 1.12 | 0.99 | 0.58 | 1.00 | 0.05 | 0.09 | 0.10 | 0.25 | 0.03 | 0.00 |
| 1997 | 0.16 | 0.32 | 0.90 | 4.72 | 3.96 | 0.94 | 0.40 | 0.16 | 0.10 | 0.05 | 0.02 | 0.02 | 0.02 | 0.00 |
| 1998 | 0.01 | 0.11 | 1.64 | 2.33 | 2.53 | 1.23 | 0.71 | 0.31 | 0.08 | 0.07 | 0.04 | 0.03 | 0.05 | 0.03 |
| 1999 | 0.57 | 0.75 | 3.71 | 0.93 | 1.25 | 1.64 | 0.57 | 0.17 | 0.02 | 0.02 | 0.02 | 0.00 | 0.00 | 0.02 |
| 2000 | 0.00 | 0.38 | 2.02 | 2.54 | 0.61 | 0.84 | 0.53 | 0.47 | 0.07 | 0.03 | 0.01 | 0.00 | 0.01 | 0.01 |
| 2001 | 0.00 | 0.89 | 1.90 | 2.64 | 1.60 | 0.20 | 0.23 | 0.40 | 0.13 | 0.07 | 0.04 | 0.01 | 0.00 | 0.00 |
| 2002 | 0.02 | 1.05 | 2.23 | 2.97 | 3.08 | 2.15 | 0.42 | 0.49 | 0.32 | 0.22 | 0.02 | 0.03 | 0.00 | 0.00 |
| 2003 | 0.01 | 0.05 | 9.62 | 5.06 | 2.94 | 1.34 | 0.77 | 0.21 | 0.05 | 0.10 | 0.02 | 0.03 | 0.00 | 0.00 |
| 2004 | 0.01 | 0.91 | 1.38 | 9.39 | 6.04 | 4.35 | 1.48 | 0.81 | 0.17 | 0.16 | 0.12 | 0.06 | 0.02 | 0.00 |
| 2005 | 0.00 | 0.26 | 4.32 | 2.39 | 7.42 | 4.66 | 2.31 | 0.86 | 0.44 | 0.12 | 0.05 | 0.08 | 0.03 | 0.00 |
| 2006 | 0.01 | 0.00 | 2.18 | 6.69 | 1.98 | 8.91 | 3.52 | 1.21 | 0.29 | 0.25 | 0.03 | 0.04 | 0.04 | 0.00 |
| 2007 | 0.00 | 0.06 | 0.31 | 1.73 | 3.22 | 0.81 | 1.62 | 0.70 | 0.29 | 0.16 | 0.11 | 0.08 | 0.02 | 0.00 |
| 2008 | 0.01 | 0.08 | 2.25 | 1.79 | 2.85 | 4.01 | 0.61 | 0.78 | 0.34 | 0.15 | 0.09 | 0.13 | 0.04 | 0.02 |
| 2009 | 0.01 | 0.21 | 2.43 | 1.80 | 0.68 | 0.91 | 0.84 | 0.12 | 0.26 | 0.15 | 0.03 | 0.04 | 0.00 | 0.02 |
| 2010 | 0.00 | 0.07 | 1.23 | 4.99 | 2.49 | 0.63 | 0.60 | 0.48 | 0.07 | 0.13 | 0.07 | 0.07 | 0.07 | 0.02 |
| 2011 | 0.00 | 0.15 | 3.83 | 4.20 | 3.06 | 1.15 | 0.41 | 0.39 | 0.44 | 0.17 | 0.10 | 0.09 | 0.06 | 0.05 |
| 2012 | 0.02 | 0.02 | 1.75 | 12.04 | 6.86 | 2.75 | 0.62 | 0.17 | 0.38 | 0.50 | 0.13 | 0.12 | 0.06 | 0.08 |
| 2013 | 0.01 | 0.12 | 4.27 | 7.43 | 6.78 | 4.65 | 2.57 | 1.12 | 0.30 | 0.44 | 0.36 | 0.26 | 0.13 | 0.01 |
| 2014 | 0.01 | 0.03 | 0.39 | 3.84 | 3.78 | 2.04 | 0.86 | 0.42 | 0.15 | 0.11 | 0.18 | 0.18 | 0.07 | 0.09 |
| 2015 | 0.06 | 0.04 | 1.07 | 1.90 | 3.16 | 1.72 | 0.81 | 0.72 | 0.68 | 0.45 | 0.26 | 0.23 | 0.21 | 0.15 |

Table 8.6. Saithe in division Va. Commercial catch-at-age residuals $\log (\mathrm{obs} / \mathrm{fit})$.

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | -0.79 | -0.60 | 0.25 | 0.16 | -0.08 | 0.22 | -0.05 | 0.22 | -0.30 | -0.41 | -0.70 | -0.04 |
| 1981 | -0.51 | -0.26 | -0.56 | 0.36 | -0.16 | 0.06 | 0.12 | 0.28 | -0.85 | 0.93 | 1.19 | 1.88 |
| 1982 | 0.84 | -0.25 | 0.13 | -0.37 | 0.24 | -0.03 | 0.37 | -0.36 | -1.20 | -1.15 | -0.53 | -0.10 |
| 1983 | -2.48 | 0.94 | -0.67 | 0.11 | 0.46 | 0.20 | -0.01 | -0.80 | -2.43 | -2.79 | -5.17 | -3.83 |
| 1984 | -4.23 | -1.61 | -1.29 | 0.22 | 0.47 | 0.79 | -0.45 | 0.43 | 0.96 | 1.03 | 3.47 | 4.90 |
| 1985 | -0.28 | 1.23 | 0.58 | 0.08 | 0.27 | -0.19 | -1.16 | -0.75 | -1.41 | -3.07 | 0.65 | 2.46 |
| 1986 | 2.28 | -0.72 | 0.24 | -0.32 | -0.09 | 0.10 | -0.41 | 0.93 | -1.09 | -1.39 | -1.83 | -1.75 |
| 1987 | -0.99 | 0.16 | 0.32 | 0.22 | 0.09 | 0.07 | 0.06 | -0.18 | -0.08 | -2.91 | -1.91 | -0.24 |
| 1988 | 0.92 | -0.31 | 0.06 | 0.04 | -0.14 | 0.20 | 0.79 | -0.66 | 0.48 | -1.69 | -3.27 | -6.88 |
| 1989 | -0.84 | 0.61 | 0.01 | 0.23 | -0.58 | 0.04 | 0.27 | -0.18 | 0.71 | 0.37 | -1.13 | -3.73 |
| 1990 | -1.78 | -0.54 | 0.07 | 0.01 | 0.34 | 0.06 | 0.11 | -0.37 | 0.07 | -0.78 | 0.16 | -1.61 |
| 1991 | -1.96 | -1.11 | 0.06 | 0.33 | 0.03 | -0.08 | 0.01 | 0.72 | 0.23 | -1.38 | -3.92 | -3.95 |
| 1992 | -0.22 | 0.58 | 1.05 | 0.43 | 0.06 | -0.83 | -0.25 | -0.40 | -0.28 | -1.17 | 0.48 | -0.88 |
| 1993 | 0.98 | -0.16 | -0.33 | -0.14 | -0.20 | -0.08 | 0.40 | 0.04 | 0.33 | 0.75 | 0.65 | -1.28 |
| 1994 | 1.09 | 0.98 | -0.12 | -0.69 | -0.43 | -0.46 | 0.20 | 0.41 | 0.51 | 0.80 | 1.87 | 1.79 |
| 1995 | 1.59 | 0.28 | 0.10 | -0.02 | 0.04 | -0.08 | -0.20 | -0.19 | -0.26 | -0.87 | -0.68 | -1.83 |
| 1996 | 1.45 | 0.17 | -0.12 | -0.51 | -0.34 | 0.19 | 0.27 | 0.03 | 0.39 | -0.15 | 1.33 | 2.41 |
| 1997 | 0.25 | 0.33 | -0.30 | 0.08 | -0.03 | 0.29 | -0.10 | -0.40 | -0.30 | 0.37 | -1.12 | 0.20 |
| 1998 | -0.36 | -0.07 | -0.47 | -0.76 | 0.18 | 0.00 | 0.84 | 0.70 | 1.24 | 1.15 | 1.57 | 0.84 |
| 1999 | 0.38 | 0.03 | 0.00 | 0.08 | 0.00 | -0.20 | -0.35 | 0.33 | -0.65 | -0.57 | 0.31 | 0.17 |
| 2000 | -0.07 | -0.20 | 0.11 | 0.07 | -0.1 | 0.48 | -0.49 | -0.27 | -0.22 | -0.93 | -0.09 | -1.12 |
| 2001 | -0.09 | 0.23 | -0.27 | -0.14 | -0.03 | -0.18 | 0.39 | 0.06 | 0.13 | 0.04 | 0.37 | 1.00 |
| 2002 | -0.62 | -0.08 | 0.18 | 0.36 | -0.17 | 0.12 | -0.26 | -0.35 | -0.09 | -1.20 | -0.10 | -0.36 |
| 2003 | 0.40 | -0.28 | 0.42 | -0.02 | -0.02 | -0.61 | 0.04 | -0.20 | 0.04 | -1.24 | 0.24 | 1.21 |
| 2004 | -0.16 | -0.39 | -0.11 | 0.27 | -0.10 | 0.28 | 0.61 | 0.41 | -0.05 | -0.47 | 0.75 | -0.26 |
| 2005 | -0.39 | -0.32 | -0.31 | 0.44 | 0.34 | -0.23 | -0.13 | -0.01 | 0.31 | -0.07 | -0.31 | -0.43 |
| 2006 | -0.69 | -0.16 | -0.35 | -0.03 | 0.56 | 0.07 | -0.33 | -0.10 | 0.76 | 0.95 | 1.05 | 0.12 |
| 2007 | 0.79 | 0.22 | 0.14 | 0.25 | -0.13 | -0.04 | -0.40 | -0.48 | -0.82 | 0.30 | 0.25 | -0.26 |
| 2008 | 0.11 | 0.36 | 0.18 | 0.23 | -0.11 | -0.23 | -0.24 | -0.30 | -0.60 | 0.10 | 2.71 | 1.70 |
| 2009 | 0.75 | 0.49 | -0.11 | -0.27 | -0.16 | 0.29 | -0.34 | -0.12 | 0.10 | 0.42 | 1.08 | 2.54 |
| 2010 | 0.46 | 0.28 | 0.11 | -0.46 | 0.03 | -0.12 | 0.47 | -0.43 | 0.07 | -0.06 | 0.84 | 1.18 |
| 2011 | 0.14 | -0.06 | -0.05 | -0.30 | -0.04 | 0.29 | 0.14 | 0.36 | -0.24 | -0.01 | 0.48 | 1.41 |
| 2012 | -0.57 | -0.37 | -0.27 | -0.20 | -0.24 | 0.06 | 0.66 | 0.55 | 1.38 | 0.12 | 0.55 | 0.91 |
| 2013 | -0.50 | -0.04 | 0.33 | -0.02 | -0.19 | -0.11 | 0.28 | 0.12 | -0.36 | 0.35 | -0.53 | 0.01 |
| 2014 | 0.16 | -0.20 | 0.50 | -0.04 | -0.22 | -0.37 | -0.24 | 0.45 | 0.28 | -0.38 | 1.99 | 0.16 |

Table 8.7. Saithe in division Va. Survey catch-at-age residuals $\log (o b s / f i t)$.

|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | -0.43 | -1.53 | -0.46 | 0.55 | 0.20 | 0.36 | -0.17 | 0.83 | -1.01 |
| 1986 | 0.78 | -0.61 | -0.69 | -0.80 | -0.50 | -0.35 | -0.47 | -0.68 | -0.39 |
| 1987 | -0.62 | 0.87 | 0.73 | 0.74 | 0.44 | 1.13 | 0.74 | 0.53 | 0.25 |
| 1988 | -0.35 | -2.15 | -1.48 | -0.96 | -0.29 | -0.46 | -0.39 | -1.31 | -0.59 |
| 1989 | 1.95 | 0.86 | -0.05 | -0.34 | -0.60 | 0.50 | 0.33 | 0.35 | -5.74 |
| 1990 | -0.12 | 0.36 | 0.45 | 0.33 | 0.91 | 0.48 | 0.88 | 0.65 | -0.47 |
| 1991 | 0.15 | -0.28 | -0.27 | -0.36 | -1.18 | -0.62 | -1.45 | -3.10 | -2.26 |
| 1992 | -0.64 | 0.03 | 0.74 | 1.23 | 0.46 | 0.66 | -0.02 | -0.70 | -1.15 |
| 1993 | 2.00 | 2.63 | 0.31 | 1.07 | 0.80 | 1.02 | 0.44 | 1.69 | 0.97 |
| 1994 | 0.88 | -0.44 | -0.09 | 0.29 | 0.18 | -0.12 | 0.84 | 1.31 | 2.32 |
| 1995 | 0.43 | 0.12 | -0.56 | -1.48 | -1.24 | -0.96 | -1.03 | -0.21 | -0.09 |
| 1996 | -0.62 | -1.29 | 0.24 | -0.41 | -0.09 | 0.52 | 1.34 | -0.87 | 0.00 |
| 1997 | 1.22 | -0.13 | 0.70 | 0.44 | -0.07 | -0.32 | -0.04 | -0.47 | -0.15 |
| 1998 | -1.50 | 1.36 | 0.36 | 0.11 | -0.41 | 0.34 | 0.23 | -0.06 | -0.18 |
| 1999 | 0.72 | 0.85 | 0.06 | -0.24 | 0.09 | -0.64 | -0.57 | -2.24 | -1.04 |
| 2000 | -0.72 | 0.10 | -0.23 | -0.30 | -0.21 | -0.55 | -0.07 | -0.87 | -1.14 |
| 2001 | 0.10 | -0.62 | -0.22 | -0.64 | -1.11 | -1.04 | -0.08 | -0.84 | -0.22 |
| 2002 | 0.13 | -0.61 | -0.72 | 0.08 | 0.18 | 0.42 | 0.60 | 0.34 | 0.36 |
| 2003 | -2.22 | 0.95 | -0.28 | -0.61 | -0.41 | -0.34 | 0.38 | -1.36 | -0.40 |
| 2004 | -0.05 | -0.12 | 0.31 | 0.07 | 0.34 | 0.37 | 0.45 | 0.80 | 0.59 |
| 2005 | -0.87 | 0.00 | -0.06 | 0.26 | 0.33 | 0.29 | 0.40 | 0.25 | 0.83 |
| 2006 | -6.50 | -0.16 | -0.07 | -0.03 | 1.07 | 0.72 | 0.23 | -0.33 | 0.06 |
| 2007 | -2.08 | -1.51 | -1.01 | -0.67 | -0.48 | -0.20 | -0.47 | -0.90 | -0.54 |
| 2008 | -2.23 | 0.39 | -0.03 | -0.17 | 0.19 | -0.09 | -0.36 | -0.77 | -1.19 |
| 2009 | -1.10 | 0.00 | -0.45 | -0.90 | -0.89 | -0.90 | -1.27 | -1.08 | -1.22 |
| 2010 | -2.62 | -0.81 | 0.25 | 0.18 | -0.39 | -0.66 | -0.85 | -1.35 | -1.41 |
| 2011 | -1.60 | 0.30 | 0.04 | -0.10 | -0.17 | -0.21 | -0.50 | -0.45 | 0.12 |
| 2012 | -3.77 | -0.50 | 1.04 | 0.82 | 0.31 | -0.24 | -0.60 | -0.05 | 0.11 |
| 2013 | -0.95 | 0.71 | 0.60 | 0.56 | 0.88 | 0.87 | 1.09 | 0.54 | 0.52 |
| 2014 | -2.84 | -1.35 | -0.03 | 0.00 | -0.32 | -0.45 | -0.62 | -0.80 | -0.24 |
| 2015 | -2.74 | -0.45 | -0.10 | -0.09 | -0.43 | -0.80 | -0.06 | 0.38 | 0.85 |

Table 8.8. Saithe in division Va. Main population estimates. The recruitment column is aligned so that the 2000 cohort is shown in the year 2000, but that cohort size is the estimated $\mathbf{N}$ at age $\mathbf{3}$ in 2003.

|  | B4+ | SSB | COHORT | Y | F4-9 | HR |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 312 | 122 | 32 | 58 | 0.29 | 19\% |
| 1981 | 304 | 130 | 42 | 59 | 0.26 | 19\% |
| 1982 | 294 | 148 | 35 | 69 | 0.30 | 23\% |
| 1983 | 270 | 147 | 67 | 58 | 0.24 | 22\% |
| 1984 | 287 | 149 | 91 | 63 | 0.23 | 22\% |
| 1985 | 299 | 139 | 50 | 57 | 0.25 | 19\% |
| 1986 | 318 | 137 | 32 | 65 | 0.28 | 20\% |
| 1987 | 335 | 128 | 21 | 81 | 0.35 | 24\% |
| 1988 | 415 | 125 | 29 | 77 | 0.32 | 19\% |
| 1989 | 397 | 127 | 15 | 82 | 0.31 | 21\% |
| 1990 | 377 | 134 | 20 | 98 | 0.35 | 26\% |
| 1991 | 336 | 143 | 18 | 102 | 0.37 | 30\% |
| 1992 | 288 | 135 | 30 | 80 | 0.37 | 28\% |
| 1993 | 230 | 112 | 25 | 72 | 0.40 | 31\% |
| 1994 | 187 | 93 | 17 | 64 | 0.45 | 34\% |
| 1995 | 152 | 70 | 9 | 49 | 0.46 | 32\% |
| 1996 | 148 | 61 | 30 | 40 | 0.41 | 27\% |
| 1997 | 155 | 62 | 31 | 37 | 0.37 | 24\% |
| 1998 | 153 | 68 | 53 | 32 | 0.30 | 21\% |
| 1999 | 131 | 72 | 63 | 31 | 0.31 | 24\% |
| 2000 | 141 | 74 | 72 | 33 | 0.33 | 23\% |
| 2001 | 161 | 80 | 26 | 32 | 0.28 | 20\% |
| 2002 | 217 | 96 | 72 | 42 | 0.31 | 19\% |
| 2003 | 276 | 118 | 42 | 52 | 0.30 | 19\% |
| 2004 | 316 | 137 | 19 | 65 | 0.26 | 20\% |
| 2005 | 282 | 147 | 27 | 69 | 0.29 | 25\% |
| 2006 | 307 | 156 | 41 | 76 | 0.31 | 25\% |
| 2007 | 278 | 152 | 41 | 64 | 0.28 | 23\% |
| 2008 | 248 | 149 | 50 | 70 | 0.32 | 28\% |
| 2009 | 224 | 137 | 45 | 61 | 0.30 | 27\% |
| 2010 | 227 | 127 | 39 | 54 | 0.27 | 24\% |
| 2011 | 239 | 122 | 21 | 51 | 0.24 | 21\% |
| 2012 | 253 | 122 | 26 | 52 | 0.23 | 20\% |
| 2013 | 268 | 128 | 32 | 58 | 0.25 | 22\% |
| 2014 | 265 | 132 | 33 | 46 | 0.19 | 18\% |
| 2015 | 255 | 139 | 33 | 57 | 0.26 | 22\% |

Table 8.9. Saithe in division Va. Stock in numbers.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 32.2 | 24.6 | 28.2 | 46.8 | 30.9 | 10.3 | 8.1 | 3.7 | 1.3 | 0.7 | 0.7 | 0.5 | 0.3 | 0.1 |
| 1981 | 47.9 | 26.4 | 20.2 | 22.7 | 35.2 | 21.2 | 6.3 | 4.6 | 2.0 | 0.7 | 0.4 | 0.4 | 0.3 | 0.2 |
| 1982 | 62.4 | 39.3 | 21.6 | 16.3 | 17.2 | 24.6 | 13.3 | 3.7 | 2.6 | 1.1 | 0.4 | 0.2 | 0.2 | 0.2 |
| 1983 | 52.8 | 51.1 | 32.1 | 17.4 | 12.2 | 11.8 | 14.8 | 7.5 | 1.9 | 1.4 | 0.6 | 0.2 | 0.1 | 0.1 |
| 1984 | 99.7 | 43.2 | 41.8 | 26.0 | 13.3 | 8.6 | 7.5 | 9.0 | 4.3 | 1.1 | 0.8 | 0.4 | 0.1 | 0.1 |
| 1985 | 136.4 | 81.6 | 35.4 | 33.8 | 19.9 | 9.4 | 5.6 | 4.6 | 5.2 | 2.5 | 0.7 | 0.5 | 0.2 | 0.1 |
| 1986 | 75.3 | 111.7 | 66.8 | 28.6 | 25.8 | 14.0 | 6.0 | 3.4 | 2.6 | 3.0 | 1.4 | 0.4 | 0.3 | 0.1 |
| 1987 | 47.6 | 61.7 | 91.4 | 53.9 | 21.6 | 17.8 | 8.7 | 3.5 | 1.8 | 1.5 | 1.6 | 0.8 | 0.2 | 0.2 |
| 1988 | 31.0 | 39.0 | 50.5 | 73.4 | 39.8 | 14.3 | 10.2 | 4.6 | 1.7 | 0.9 | 0.7 | 0.9 | 0.4 | 0.1 |
| 1989 | 44.0 | 25.4 | 31.9 | 40.6 | 54.7 | 26.8 | 8.5 | 5.6 | 2.3 | 0.9 | 0.5 | 0.4 | 0.5 | 0.2 |
| 1990 | 22.1 | 36.0 | 20.8 | 25.7 | 30.4 | 37.2 | 16.2 | 4.7 | 2.9 | 1.2 | 0.5 | 0.3 | 0.2 | 0.3 |
| 1991 | 29.5 | 18.1 | 29.5 | 16.7 | 19.0 | 20.2 | 31.4 | 8.6 | 2.3 | 1.5 | 0.6 | 0.2 | 0.1 | 0.1 |
| 1992 | 26.3 | 24.2 | 14.8 | 23.6 | 12.3 | 12.4 | 11.3 | 16.2 | 4.0 | 1.1 | 0.7 | 0.3 | 0.1 | 0.1 |
| 1993 | 44.3 | 21.5 | 19.8 | 11.9 | 17.4 | 8.0 | 7.0 | 5.9 | 7.7 | 2.0 | 0.5 | 0.4 | 0.2 | 0.1 |
| 1994 | 38.0 | 36.3 | 17.6 | 15.9 | 8.7 | 11.2 | 4.4 | 3.5 | 2.7 | 3.6 | 0.9 | 0.3 | 0.2 | 0.1 |
| 1995 | 25.0 | 31.1 | 29.7 | 14.1 | 11.4 | 5.4 | 5.8 | 2.1 | 1.5 | 1.2 | 1.5 | 0.4 | 0.1 | 0.1 |
| 1996 | 12.8 | 20.5 | 25.5 | 23.7 | 10.1 | 7.0 | 2.8 | 2.7 | 0.8 | 0.6 | 0.5 | 0.7 | 0.2 | 0.1 |
| 1997 | 44.9 | 10.5 | 16.8 | 20.4 | 17.3 | 6.5 | 3.8 | 1.4 | 1.2 | 0.4 | 0.3 | 0.2 | 0.4 | 0.1 |
| 1998 | 46.2 | 36.8 | 8.6 | 13.3 | 14.5 | 11.2 | 3.9 | 2.1 | 0.7 | 0.6 | 0.2 | 0.1 | 0.1 | 0.2 |
| 1999 | 79.7 | 37.8 | 30.1 | 6.8 | 9.6 | 9.8 | 7.1 | 2.3 | 1.1 | 0.3 | 0.3 | 0.1 | 0.1 | 0.1 |
| 2000 | 93.4 | 65.2 | 31.0 | 23.9 | 5.0 | 6.5 | 6.1 | 4.1 | 1.2 | 0.6 | 0.2 | 0.1 | 0.0 | 0.0 |
| 2001 | 106.8 | 76.5 | 53.4 | 24.6 | 17.2 | 3.3 | 4.0 | 3.5 | 2.1 | 0.6 | 0.3 | 0.1 | 0.1 | 0.0 |
| 2002 | 38.1 | 87.5 | 62.6 | 42.6 | 18.0 | 11.8 | 2.1 | 2.4 | 1.9 | 1.1 | 0.3 | 0.1 | 0.0 | 0.0 |
| 2003 | 107.7 | 31.2 | 71.6 | 49.8 | 30.9 | 12.2 | 7.4 | 1.2 | 1.3 | 1.0 | 0.6 | 0.2 | 0.1 | 0.0 |
| 2004 | 62.2 | 88.2 | 25.5 | 57.0 | 36.2 | 20.9 | 7.7 | 4.3 | 0.7 | 0.6 | 0.5 | 0.3 | 0.1 | 0.0 |
| 2005 | 28.1 | 50.9 | 72.2 | 19.9 | 38.0 | 22.8 | 13.0 | 4.8 | 2.7 | 0.4 | 0.4 | 0.3 | 0.2 | 0.0 |
| 2006 | 40.4 | 23.0 | 41.7 | 56.2 | 13.0 | 23.3 | 13.8 | 7.9 | 3.0 | 1.6 | 0.2 | 0.2 | 0.1 | 0.1 |
| 2007 | 61.0 | 33.1 | 18.8 | 32.4 | 36.2 | 7.9 | 13.8 | 8.2 | 4.8 | 1.7 | 0.9 | 0.1 | 0.1 | 0.1 |
| 2008 | 61.1 | 50.0 | 27.1 | 14.7 | 21.2 | 22.3 | 4.8 | 8.5 | 5.1 | 2.8 | 0.9 | 0.4 | 0.1 | 0.0 |
| 2009 | 74.2 | 50.0 | 40.9 | 21.0 | 9.3 | 12.6 | 13.0 | 2.8 | 5.0 | 2.9 | 1.5 | 0.5 | 0.2 | 0.0 |
| 2010 | 66.6 | 60.8 | 41.0 | 31.8 | 13.6 | 5.6 | 7.5 | 7.8 | 1.7 | 2.9 | 1.6 | 0.7 | 0.2 | 0.1 |
| 2011 | 58.4 | 54.5 | 49.7 | 32.0 | 21.1 | 8.5 | 3.5 | 4.7 | 4.9 | 1.0 | 1.6 | 0.8 | 0.4 | 0.1 |
| 2012 | 30.6 | 47.8 | 44.6 | 39.0 | 21.7 | 13.6 | 5.4 | 2.2 | 3.0 | 3.0 | 0.6 | 0.9 | 0.5 | 0.2 |
| 2013 | 39.3 | 25.1 | 39.2 | 35.1 | 26.7 | 14.1 | 8.7 | 3.5 | 1.4 | 1.9 | 1.8 | 0.3 | 0.5 | 0.3 |
| 2014 | 47.6 | 32.1 | 20.5 | 30.7 | 23.6 | 17.0 | 8.9 | 5.5 | 2.2 | 0.9 | 1.1 | 1.0 | 0.2 | 0.3 |
| 2015 | 48.8 | 39.0 | 26.3 | 16.2 | 21.6 | 16.0 | 11.4 | 6.0 | 3.7 | 1.5 | 0.6 | 0.7 | 0.6 | 0.1 |

Table 8.10. Saithe in division Va. Fishing mortality rate.

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.02 | 0.09 | 0.18 | 0.30 | 0.36 | 0.44 | 0.41 | 0.44 | 0.36 | 0.36 | 0.36 | 0.36 |
| 1981 | 0.01 | 0.08 | 0.16 | 0.26 | 0.32 | 0.39 | 0.36 | 0.39 | 0.32 | 0.32 | 0.32 | 0.32 |
| 1982 | 0.02 | 0.09 | 0.18 | 0.30 | 0.37 | 0.45 | 0.42 | 0.45 | 0.37 | 0.37 | 0.37 | 0.37 |
| 1983 | 0.01 | 0.07 | 0.15 | 0.24 | 0.30 | 0.36 | 0.34 | 0.36 | 0.30 | 0.30 | 0.30 | 0.30 |
| 1984 | 0.01 | 0.07 | 0.14 | 0.23 | 0.29 | 0.34 | 0.32 | 0.34 | 0.28 | 0.28 | 0.28 | 0.28 |
| 1985 | 0.01 | 0.07 | 0.15 | 0.25 | 0.30 | 0.37 | 0.34 | 0.37 | 0.30 | 0.30 | 0.30 | 0.30 |
| 1986 | 0.02 | 0.08 | 0.17 | 0.28 | 0.35 | 0.42 | 0.39 | 0.42 | 0.34 | 0.34 | 0.34 | 0.34 |
| 1987 | 0.02 | 0.10 | 0.21 | 0.35 | 0.43 | 0.52 | 0.49 | 0.52 | 0.43 | 0.43 | 0.43 | 0.43 |
| 1988 | 0.02 | 0.09 | 0.19 | 0.32 | 0.40 | 0.48 | 0.45 | 0.48 | 0.39 | 0.39 | 0.39 | 0.39 |
| 1989 | 0.02 | 0.09 | 0.19 | 0.31 | 0.38 | 0.46 | 0.43 | 0.46 | 0.37 | 0.37 | 0.37 | 0.37 |
| 1990 | 0.02 | 0.10 | 0.21 | 0.35 | 0.43 | 0.52 | 0.48 | 0.52 | 0.43 | 0.43 | 0.43 | 0.43 |
| 1991 | 0.02 | 0.11 | 0.23 | 0.38 | 0.46 | 0.56 | 0.52 | 0.56 | 0.46 | 0.46 | 0.46 | 0.46 |
| 1992 | 0.02 | 0.11 | 0.22 | 0.37 | 0.45 | 0.55 | 0.51 | 0.55 | 0.45 | 0.45 | 0.45 | 0.45 |
| 1993 | 0.02 | 0.12 | 0.24 | 0.40 | 0.49 | 0.59 | 0.55 | 0.59 | 0.49 | 0.49 | 0.49 | 0.49 |
| 1994 | 0.03 | 0.13 | 0.27 | 0.45 | 0.56 | 0.67 | 0.63 | 0.67 | 0.55 | 0.55 | 0.55 | 0.55 |
| 1995 | 0.03 | 0.13 | 0.28 | 0.46 | 0.57 | 0.69 | 0.64 | 0.69 | 0.56 | 0.56 | 0.56 | 0.56 |
| 1996 | 0.02 | 0.12 | 0.25 | 0.41 | 0.50 | 0.60 | 0.56 | 0.60 | 0.49 | 0.49 | 0.49 | 0.49 |
| 1997 | 0.04 | 0.14 | 0.23 | 0.31 | 0.42 | 0.53 | 0.57 | 0.55 | 0.56 | 0.56 | 0.56 | 0.56 |
| 1998 | 0.03 | 0.12 | 0.19 | 0.26 | 0.34 | 0.43 | 0.46 | 0.45 | 0.46 | 0.46 | 0.46 | 0.46 |
| 1999 | 0.03 | 0.12 | 0.20 | 0.27 | 0.36 | 0.45 | 0.49 | 0.47 | 0.48 | 0.48 | 0.48 | 0.48 |
| 2000 | 0.03 | 0.13 | 0.21 | 0.28 | 0.38 | 0.47 | 0.51 | 0.50 | 0.51 | 0.51 | 0.51 | 0.51 |
| 2001 | 0.03 | 0.11 | 0.18 | 0.24 | 0.32 | 0.40 | 0.43 | 0.42 | 0.43 | 0.43 | 0.43 | 0.43 |
| 2002 | 0.03 | 0.12 | 0.19 | 0.26 | 0.35 | 0.44 | 0.47 | 0.46 | 0.47 | 0.47 | 0.47 | 0.47 |
| 2003 | 0.03 | 0.12 | 0.19 | 0.26 | 0.34 | 0.43 | 0.47 | 0.45 | 0.46 | 0.46 | 0.46 | 0.46 |
| 2004 | 0.05 | 0.21 | 0.26 | 0.28 | 0.27 | 0.26 | 0.30 | 0.37 | 0.43 | 0.43 | 0.43 | 0.43 |
| 2005 | 0.05 | 0.22 | 0.29 | 0.31 | 0.29 | 0.29 | 0.32 | 0.40 | 0.47 | 0.47 | 0.47 | 0.47 |
| 2006 | 0.05 | 0.24 | 0.31 | 0.33 | 0.31 | 0.31 | 0.35 | 0.43 | 0.51 | 0.51 | 0.51 | 0.51 |
| 2007 | 0.05 | 0.22 | 0.28 | 0.30 | 0.29 | 0.28 | 0.32 | 0.39 | 0.46 | 0.46 | 0.46 | 0.46 |
| 2008 | 0.06 | 0.25 | 0.32 | 0.34 | 0.33 | 0.33 | 0.37 | 0.45 | 0.53 | 0.53 | 0.53 | 0.53 |
| 2009 | 0.05 | 0.24 | 0.30 | 0.32 | 0.31 | 0.30 | 0.34 | 0.42 | 0.50 | 0.50 | 0.50 | 0.50 |
| 2010 | 0.05 | 0.21 | 0.27 | 0.28 | 0.27 | 0.27 | 0.30 | 0.37 | 0.44 | 0.44 | 0.44 | 0.44 |
| 2011 | 0.04 | 0.19 | 0.24 | 0.26 | 0.25 | 0.24 | 0.27 | 0.34 | 0.40 | 0.40 | 0.40 | 0.40 |
| 2012 | 0.04 | 0.18 | 0.23 | 0.25 | 0.23 | 0.23 | 0.26 | 0.32 | 0.38 | 0.38 | 0.38 | 0.38 |
| 2013 | 0.04 | 0.20 | 0.25 | 0.27 | 0.25 | 0.25 | 0.28 | 0.35 | 0.41 | 0.41 | 0.41 | 0.41 |
| 2014 | 0.03 | 0.15 | 0.19 | 0.20 | 0.20 | 0.19 | 0.22 | 0.27 | 0.32 | 0.32 | 0.32 | 0.32 |
| 2015 | 0.05 | 0.20 | 0.26 | 0.28 | 0.26 | 0.26 | 0.29 | 0.36 | 0.43 | 0.43 | 0.43 | 0.43 |

Table 8.11. Saithe in division Va. Input values for short term projections. Same weights are used for catch weights and stock weights.

| 201 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| N | 26.3 | 16.2 | 21.6 | 16.0 | 11.4 | 6.0 | 3.7 | 1.5 | 0.6 | 0.7 | 0.6 | 0.1 |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| mat | 0.00 | 0.10 | 0.21 | 0.40 | 0.63 | 0.81 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 0 | 0 | 7 | 9 | 3 | 2 | 5 | 0 | 0 | 0 | 0 | 0 |
| W | 1.14 | 1.72 | 2.21 | 3.07 | 4.03 | 5.53 | 6.84 | 7.17 | 7.49 | 8.21 | 8.57 | 9.17 |
|  | 2 | 6 | 7 | 1 | 0 | 2 | 6 | 5 | 1 | 8 | 5 | 3 |
| sel | 0.10 | 0.47 | 0.60 | 0.64 | 0.62 | 0.61 | 0.68 | 0.84 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 6 | 6 | 8 | 8 | 1 | 1 | 7 | 6 | 0 | 0 | 0 | 0 |
| pF | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| pM | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $6$ |  |  |  |  |  |  |  |  |  |  |  | 14 |
| N | 31.9 |  |  |  |  |  |  |  |  |  |  |  |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| mat | 0.00 | 0.10 | 0.21 | 0.40 | 0.63 | 0.81 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 0 | 0 | 7 | 9 | 3 | 2 | 5 | 0 | 0 | 0 | 0 | 0 |
| W | 1.14 | 1.72 | 2.21 | 3.07 | 4.03 | 5.53 | 6.84 | 7.17 | 7.49 | 8.21 | 8.57 | 9.17 |
|  | 2 | 6 | 7 | 1 | 0 | 2 | 6 | 5 | 1 | 8 | 5 | 3 |
| sel | 0.10 | 0.47 | $0.60$ | $0.64$ | $0.62$ | $0.61$ | $0.68$ | $0.84$ | $1.00$ | $1.00$ | $1.00$ | 1.00 |
|  | 6 | 6 | 8 | $8$ | $1$ | $1$ | $7$ | $6$ | $0$ | $0$ | $0$ |  |
| pF | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| pM | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $7$ |  |  |  |  |  |  |  |  |  |  |  |  |
| N | 32.7 |  |  |  |  |  |  |  |  |  |  |  |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| mat | 0.00 | 0.10 | 0.21 | 0.40 | 0.63 | 0.81 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 0 | 0 | 7 | 9 | 3 | 2 | 5 | 0 | 0 | 0 | 0 | 0 |
| W | 1.14 | 1.72 | 2.21 | 3.07 | 4.03 | 5.53 | 6.84 | 7.17 | 7.49 | 8.21 | 8.57 | 9.17 |
|  | 2 | 6 | 7 | 1 | 0 | 2 | 6 | 5 | 1 | 8 | 5 | 3 |
| sel | 0.10 | 0.47 | 0.60 | 0.64 | 0.62 | 0.61 | 0.68 | 0.84 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 6 | 6 | 8 | 8 | 1 | 1 | 7 | 6 | 0 | 0 | 0 |  |
| pF | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| pM | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 8.12. Saithe in division Va. Output from short-term projections.

| 2015 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| B4+ | SSB | FBAR | LANDINGS |  |  |  |
| 255 | 139 | 0.26 | 57 |  |  |  |
| 2016 |  |  |  | 2017 |  |  |
| B4+ | SSB | FBAR | LANDINGS | B4+ | SSB | RATIONALE |
| 238 | 138 | 0.26 | 54 | 227 | 130 | $20 \%$ HCR |

$20 \% \mathrm{HCR}=$ average between 0.2 B4+ (current year) and last year's TAC


Figure 8.1 Saithe in Division Va. Landings by gear.
Saithe - Apr 122015
Based on landings data from the Directorate of Fisheries



Figure 8.2 Saithe in division Va. Cumulative landings in the current fishing year (left) and calendar year (right). The vertical (green line) in the left figure shows the quota for the current fishing year.


Figure 8.3 Saithe in division Va. Weight at age in the catches, as relative deviations from the mean The current year's deviation is a preliminary prediction.


Figure 8.4 Saithe in division Va. Maturity-at-age used for calculating the SSB.


Figure 8.5 Saithe in division Va. Spring survey biomass index and model fit.


Figure 8.6. Estimated selectivity patterns for the 3 periods.

Std catch residuals


Figure 8.7. Saithe in division Va. Commercial and survey catch-at-age residuals from the fitted model. Filled circles are positive log residuals and hollow circles are negative log residuals.







Figure 8.8. Saithe in division Va. Results from the fitted model and short-term forecast. The red line indicates the time of the current assessment.




Figure 8.9. Saithe in division Va. Retrospective pattern for the assessment model.


Figure 8.10. Saithe in division Va. Comparison between the default separable model (ADSEP) and alternative assessment models.

## Summary

This stock was benchmarked in 2015. A formal HCR is in place to determine the TAC for this stock since 1994. The rule has gone through amendments and revisions. The last significant change occurred in 2007, when the harvest rate multiplier upon which the TAC for the next fishing season is based was changed from $0.25-0.20$ of the reference biomass ( $B_{4+}$ - biomass of age 4 years and older) as estimated in the beginning of the assessment year. The HCR upon which the TAC is set when the SSB in the assessment year is estimated to be above $S S B_{\text {trigger }}(220 \mathrm{kt})$ is as follows:

$$
T A C_{y / y+1}=\left(0.20 * B_{4+, y}+T A C_{y-1 / y}\right) / 2
$$

The decision on the TAC each year is based on an analytical stock assessment carried out each year.

The results from the assessment show that the spawning stock $\left(S S B_{2016}\right)$ is estimated to be 464 kt . The high values estimated in recent years are higher than has been observed over the last five decades. The reference biomass $\left(B_{4+, 2016}\right)$ is estimated to be 1241 kt , and has not been so high since the late 1970's. Fishing mortality, being 0.27 in 2015, has declined significantly in recent years and is presently the lowest observed in last 6 decades. Year classes since the mid-1980s are estimated to be relatively stable but with the mean around the lower values observed in the period 1955 to 1985. Estimates of year classes 2014 and 2015 indicate that they may be larger than that observed in the recent decades.

Mean weight at age in the stock and the catches that were record low in 2006-8 have been increasing in recent years and are now around the long term mean. Spring survey weights in some of the important age groups have however declined in 2016 compared with 2015. Catch weights are hence estimated to be more pessimistic in 2016 compared to 2015.

The input in the analytical assessments are catch at age 1955-2015 and spring groundfish survey (SMB) indices at age from 1985-2016 and fall survey groundfish survey (SMH) indices at age from 1996-2015. The results from the AD-Model builder statistical Catch at Age Model (ADCAM) as was used as the final run. This framework has been the basis for the advice since 2002 (spring survey only up to the 2009 assessment, both surveys since then).

### 9.1.1 Data

The data used for assessing Icelandic cod landings, catch-at-age composition and indices from standardized bottom trawl surveys. The sampling programs i.e. log books, surveys, sampling from landings etc. have been described in previous reports.

### 9.1.1.1 Landings

Landings of Icelandic cod in 2015 are estimated to have been 230 kt of which 228 kt where taken by Icelandic fleet.

Historically the landings of bottom trawlers constituted a larger portion of the total catches than today, in some years prior to 1990 reaching $60 \%$ of the total landings. In the 1990's the landings from bottom trawlers declined significantly within a period of

5 years, and have been just above $40 \%$ of the total landings in the last decade. The share of long line has tripled over the last 20 years and is now on par with bottom trawl. The share of gill net has over the same time period declined and is now only half of what it was in the 1980's.

The landings statistics for the period 2001-2015 were re-evaluated in this year assessment. The details are described in WD27 (iCod - Some details), the revisions accounting changes within $1 \%$ of the total landings previously used.

The trend in landings in last two decades is largely a reflection of the TAC that is set for the fishing year (starting 1 September and ending 31 August). According to the HCR the catch for the fishing year $2014 / 2015$ was supposed to be capped to 218 kt . Landings of the Icelandic fleet was 221 kt . Including additional landings of some 2 kt by the foreign fleet this amounts to an approximate $2 \%$ overshoot.

The estimates of landing for the current calendar year of 246 kt is based on the remainder of the quota from the current fishing year $(2015 / 16)$ on 1. January 2016 ( 149 kt ), the catch that is expected to be taken from 1. September to 31. December $2016(95 \mathrm{kt})$ and the expected catch of the foreign fleet ( 2 kt ).

Mean annual discard of cod over the period 2001-2012 is around $1 \%$ of landings (Ólafur Pálsson et al 2013). The method used for deriving these estimates assumes that discarding only occurs as high grading.

### 9.1.1.2 Catch in numbers and weight by age\}

Catch in numbers by age: The method for deriving the catch at age is based on 20 métiers: two areas (north and south), two seasons (January-May and June-December) and four fleets (bottom trawl, longline, hooks (jiggers), gillnet and Danish seine). Until now an ad hoc method has been used when "borrowing" age-length keys and/or length distribution for métiers where catches and hence sampling have been low. The programs used are old prelude scripts that are difficult to maintain and atomize. Different approaches may have been taken in different years making the calculations non-reproducible. An equivalent algorithm is available as an $R$-script (has been used to derive the catch at age in Icelandic haddock for numerous years) and was used this year to derive the catch-at-age matrix for cod (period 2001-2015). The difference in the two algorithms (prelude vs $R$ ) pertain to the handling of métiers where the catches and hence samplings are low. In the latter case a low weight (0.01) of a combined age-length key is provided in addition the data available in each métier. This resulted in some difference in the catch at age within age groups (WD27: iCod - Some details), but effect on the key stock assessment metrics are insignificant (Figure 1).

The catch at age matrix is reasonably consistent (Table 1), with CV estimated to be approximately 0.2 for age groups 4-10 based on a Shepherd-Nicholson model.

Mean weight at age in the landings: The mean weight age in the landings (Table 2 and Figure 2) declined from 2001-2007, reaching then a historical low in many age groups. The weight at age have been increasing in recent years and are in 2015 around the average weights observed over the period from 1985 and close to the long term mean (1955-2015). The variation in the pattern of weight at age in the catches is in part a reflection of the variation in the weight in the stock as seen in the measurements from the spring survey (Table 5 and Figure 3). The latest spring survey weight measurements (in 2016) are below average in younger ages but above average in older ages.

The reference biomass ( $B_{4+}$ ) upon which the TAC in the fishing year is set (based on the HCR) is derived from population numbers and catch weights in the beginning of the assessment year. In recent years the estimates of mean weights in the landings of age groups 3-9 in the assessment years have been based on a prediction from the spring survey weight measurements in that year using the slope and the intercept from a linear relationship between survey and catch weights in preceding year. The same approach was used this year for predicting weight at age in the catches for 2016. I.e. the alpha and beta were estimated from :

$$
c W_{a, 2015}=a l p h a+\text { bet } a * s W_{a, 2015}
$$

and the catch weights for 2016 then from:

$$
c W_{a, 2016}=a l p h a+\text { beta } * s W_{a, 2016}
$$

Based on this the mean weights at age in the catches in 2016 are predicted to be at or somewhat below the average (Figure 2).

### 9.1.1.3 Surveys

Length based indices: The total biomass indices from the spring (SMB) and the fall (SMH) surveys (Figure 4) indicate that the stock biomass has been increasing substantially in recent years and is in the last 5 years among highest since the start of the spring survey in 1985. The increase in biomass is most pronounced in larger fish.

Age based indices: Abundance indices by age from the spring and the fall surveys (Tables 6 and 7). Indices of older fish are all relatively high in recent years despite the indices of these year classes when younger are low or moderate in size (Figure 5).

### 9.1.2 Assessment

The results from a statistical catch at age model (sometimes refer to as ADCAM) tuned with the spring and the fall survey have been used as the final point estimator upon which advice was based. In this framework the catch at age are modeled and the fishing mortality changes gradually over time, constrained by a random walk. The survey residuals in a given year are modeled by a multivariate normal distribution to account for potential survey "year effects".

The tuning with both the spring and the fall survey show similar diagnostics as that observed in previous years (see Tables 8, 9 and 10 and Figure 6 for the residuals). A negative residual block for spring survey indices age groups 2 to 5 in recent years may indicate that there may have been some change in catchability. The detailed result from the assessment is provided in Tables 11, 12 and the stock summary in Table 13 and Figure 7. The reference biomass is estimated to be 1241 kt in 2016 and the fishing mortality 0.27 in 2015.

Assessment based on tuning with the spring and the fall survey separately have in recent years shown that the fall survey gives a higher estimate than the spring survey. Tuning with spring survey only this year resulted in a reference biomass of 1136 kt in 2016 and a fishing mortality of 0.30 in 2015. An assessment based on the fall survey only gave reference biomass of 1291 kt in 2016 and fishing mortality of 0.26 in 2015.

The reference stock ( $B_{4+}$ ) in 2015 is now estimated to be 1254 kt compared to 1302 kt last year. The SSB in 2015 is now estimated to be 533 kt compared to 547 kt estimated last year. Fishing mortality in 2014 is now estimated 0.29 compared to 0.28 estimated
last year. Year classes 2012-2014 were estimated to be 161, 115 and 186 million in last year's assessment and are now estimated to be 165, 117 and 208 million.

### 9.1.3 HCR and reference points

The HCR upon which the TAC is set when the SSB in the assessment year is estimated to be above SSB $_{\text {trigger }}(220 \mathrm{kt}$ ) is as follows:

$$
T A C_{y / y+1}=\left(0.20 * B_{4+, y}+T A C_{y-1 / y}\right) / 2
$$

In case the SSB is estimated to be below $S S B_{\text {trigger }}=220 \mathrm{kt}$ the 0.20 multipler is reduced linearily. The $B_{4+, y}$ refers to the refence biomass (4 years and older) in the beginning of the assessment year $(y)$. The notation $y / y+1$ refers to the next fishing year (starting 1. September of the assessent year) and $y-1 / y$ to the current fishing year (ending 31. August of the assessment year). The advice for the 2016/2017 fishing season is:

$$
T A C_{y / y+1}=(0.20 * 1241+239) / 2=244 k t
$$

Although no prediction (besides catch weights in the assessment year) are needed to derive the advice, the basis as well as the calculation are provided (Table 14 and 15).

The rule was formally evaluated by ICES in 2009, but had been in place since the 2007/2008 fishing season. The evaluation showed that using the 0.20 multiplier would result in yield that was close to maximum (maximum yield when no catch stabilizer is used was estimated when applying a multiplier of around 0.22 ), while at the same time have a low probability that the stock would go below the $S S B_{\text {trigger }}$. The results were robust to numerous stock-recruitment scenarios tested, including assumption that future maximum mean recruitment would be around the mean observed since 1985 (Figure 8). All scenarios tested showed that there was very low probability that the stock would go below $B_{\text {lim }}=B_{\text {loss }}=125 k t$ (formally set in 2010) if the above rule is followed. ICES concluded that the HCR was in conformity both the ICES PA and MSY approach

Assessment errors (CV $=0.15, \rho=0.45$ (autocorrelation)) were implicitly included in the HCR evaluations. These errors were estimated from empirical retrospective pattern in the estimates of the reference biomass since the earliest available assessment in the 1970's. The distribution of the realized harvest rate when the HCR is followed showed that the $90 \%$ expected range are within a harvest rate of $0.15-0.27$ (Figure 10). Hence classifying recent harvest rates above or below the 0.20 advisory multiplier is truly nonsensical. The recent realized harvest rates are within the above range.

In a ToR for the NWWG this year a request is made to derive pa-reference points for all ICES stocks for stock classification purpose. This is irrespective of which management jurisdiction a stock falls under.

The estimated CV of the spawning stock in the assessment year is around 0.08. The cv of the fishing mortality in the year before the assessment year is around 0.06 . These estimates are an underestimates of the uncertainty.

Hence, for the derivation of pa-points from lim-points was based on a $\mathrm{CV}=0.15$. The $B_{p a}$ is derived as:

$$
B_{p a}=B_{\text {lim }} e^{1.645 * c v}=125 e^{1.645 * 0.15}=160
$$

The fishing mortality reference points that lead to $B_{\text {lim }}$ was derived using the same framework as was used in the HCR evaluation (with no $B_{\text {trigger }}$, no assessment error and no autocorrelation in recruitment) is around 0.74 . Hence the $F_{p a}$ is derived as:

$$
F_{p a}=F_{l i m} e^{-1.645 * c v}=0.74 e^{-1.645 * 0.15}=0.58
$$

The $F_{p a}$ reference point is higher than the approximate fishing mortality that correspond to the 0.20 harvest rate (fishing mortality around 0.30 ) that is the basis of the advice.

### 9.1.4 On frequency of assessment

ICES is considering if advice and hence assessments can be done less frequently than annually. As a part of that a ToR for this year was to provide a 7 year peel of the Mohn's statistic based on the analytical retrospective on reference F. For the cod the mean value is 0.04 for the assessment years 2009-2015 relative to the current assessment, the annual range being 0.02-0.08 (Figure 12).

Given the criterion that candidate stocks for inter-annual assessments are those that have a Mohn's based on the average of a 7 year peel of less than 0.2 , this stock could be considered as a candidate. Counter to this is that this stock is under a management plan that includes annual setting of TAC based on a HCR. The evaluation of the HCR, including parameter settings such as assessment error (cv and autocorrelation) were set based on the CV in the assessment year. If TAC for 2-3 were to be set based on a less frequent assessment a re-evaluation of the HCR would be required because the CV in the prediction years would be significantly higher. It would most likely result in that the multiplier (harvest rate) in the HCR would be lower than at present.

In general the same argument could apply to stocks that are under the ICES MSY advice. Here, the derived $F_{m s y}$ is dependent on the assumption of assessment errors. The current values used are probably in all cases based on annual assessment errors. If advice were to be set multi-annually, the input assessment error would need to be set significantly higher when estimating of the advisory $F_{m s y}$, particularly in stocks were no direct measurements is available of the future recruits. The effect of setting higher cv's would be most pronounced in stocks where the derived $F_{m s y}$ were bounded by precautionary consideration (low probability of going below $B_{\text {lim }}$ ). The net effect is that a multi-annual advice needs to be based on a lower advisory fishing mortality than used if the advice were annual.

### 9.2 Tables

Table 1. Icelandic cod in Division 5.a. Estimated catch in numbers by year and age in millions of fish in 1955-2015

| YEAR | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 4.790 | 25.164 | 46.566 | 28.287 | 10.541 | 5.224 | 2.467 | 25.182 | 2.101 | 1.202 | 1.668 | 0.665 |
| 1956 | 6.709 | 17.265 | 31.030 | 27.793 | 14.389 | 4.261 | 3.429 | 2.128 | 16.820 | 1.552 | 1.522 | 1.545 |
| 1957 | 13.240 | 21.278 | 17.515 | 24.569 | 17.634 | 12.296 | 3.568 | 2.169 | 1.171 | 6.822 | 0.512 | 1.089 |
| 1958 | 25.237 | 30.742 | 14.298 | 10.859 | 15.997 | 15.822 | 12.021 | 2.003 | 2.125 | 0.771 | 3.508 | 0.723 |
| 1959 | 18.394 | 37.650 | 23.901 | 7.682 | 5.883 | 8.791 | 13.003 | 7.683 | 0.914 | 0.990 | 0.218 | 1.287 |
| 1960 | 14.830 | 28.642 | 27.968 | 14.120 | 8.387 | 6.089 | 6.393 | 11.600 | 3.526 | 0.692 | 0.183 | 0.510 |
| 1961 | 16.507 | 21.808 | 19.488 | 15.034 | 7.900 | 6.925 | 3.969 | 3.211 | 6.756 | 1.202 | 0.089 | 0.425 |
| 1962 | 13.514 | 28.526 | 18.924 | 14.650 | 12.045 | 4.276 | 8.809 | 2.664 | 1.883 | 2.988 | 0.405 | 0.324 |
| 1963 | 18.507 | 28.466 | 19.664 | 11.314 | 15.682 | 7.704 | 2.724 | 6.508 | 1.657 | 1.030 | 1.372 | 0.246 |
| 1964 | 19.287 | 28.845 | 18.712 | 11.620 | 7.936 | 18.032 | 5.040 | 1.437 | 2.670 | 0.655 | 0.370 | 1.025 |
| 1965 | 21.658 | 29.586 | 24.783 | 11.706 | 9.334 | 6.394 | 11.122 | 1.477 | 0.823 | 0.489 | 0.118 | 0.489 |
| 1966 | 17.910 | 30.649 | 20.006 | 13.872 | 5.942 | 7.586 | 2.320 | 5.583 | 0.407 | 0.363 | 0.299 | 0.311 |
| 1967 | 25.945 | 27.941 | 24.322 | 11.320 | 8.751 | 2.595 | 5.490 | 1.392 | 1.998 | 0.109 | 0.030 | 0.106 |
| 1968 | 11.933 | 47.311 | 22.344 | 16.277 | 15.590 | 7.059 | 1.571 | 2.506 | 0.512 | 0.659 | 0.047 | 0.098 |
| 1969 | 11.149 | 23.925 | 45.445 | 17.397 | 12.559 | 14.811 | 1.590 | 0.475 | 0.340 | 0.064 | 0.024 | 0.021 |
| 1970 | 9.876 | 47.210 | 23.607 | 25.451 | 15.196 | 12.261 | 14.469 | 0.567 | 0.207 | 0.147 | 0.035 | 0.050 |
| 1971 | 13.060 | 35.856 | 45.577 | 21.135 | 17.340 | 10.924 | 6.001 | 4.210 | 0.237 | 0.069 | 0.038 | 0.020 |
| 1972 | 8.973 | 29.574 | 30.918 | 22.855 | 11.097 | 9.784 | 10.538 | 3.938 | 1.242 | 0.119 | 0.031 | 0.001 |
| 1973 | 36.538 | 25.542 | 27.391 | 17.045 | 12.721 | 3.685 | 4.718 | 5.809 | 1.134 | 0.282 | 0.007 | 0.001 |
| 1974 | 14.846 | 61.826 | 21.824 | 14.413 | 8.974 | 6.216 | 1.647 | 2.530 | 1.765 | 0.334 | 0.062 | 0.028 |
| 1975 | 29.301 | 29.489 | 44.138 | 12.088 | 9.628 | 3.691 | 2.051 | 0.752 | 0.891 | 0.416 | 0.060 | 0.046 |
| 1976 | 23.578 | 39.790 | 21.092 | 24.395 | 5.803 | 5.343 | 1.297 | 0.633 | 0.205 | 0.155 | 0.065 | 0.029 |
| 1977 | 2.614 | 42.659 | 32.465 | 12.162 | 13.017 | 2.809 | 1.773 | 0.421 | 0.086 | 0.024 | 0.006 | 0.002 |
| 1978 | 5.999 | 16.287 | 43.931 | 17.626 | 8.729 | 4.119 | 0.978 | 0.348 | 0.119 | 0.048 | 0.015 | 0.027 |
| 1979 | 7.186 | 28.427 | 13.772 | 34.443 | 14.130 | 4.426 | 1.432 | 0.350 | 0.168 | 0.043 | 0.024 | 0.004 |
| 1980 | 4.348 | 28.530 | 32.500 | 15.119 | 27.090 | 7.847 | 2.228 | 0.646 | 0.246 | 0.099 | 0.025 | 0.004 |
| 1981 | 2.118 | 13.297 | 39.195 | 23.247 | 12.710 | 26.455 | 4.804 | 1.677 | 0.582 | 0.228 | 0.053 | 0.068 |
| 1982 | 3.285 | 20.812 | 24.462 | 28.351 | 14.012 | 7.666 | 11.517 | 1.912 | 0.327 | 0.094 | 0.043 | 0.011 |
| 1983 | 3.554 | 10.910 | 24.305 | 18.944 | 17.382 | 8.381 | 2.054 | 2.733 | 0.514 | 0.215 | 0.064 | 0.037 |
| 1984 | 6.750 | 31.553 | 19.420 | 15.326 | 8.082 | 7.336 | 2.680 | 0.512 | 0.538 | 0.195 | 0.090 | 0.036 |
| 1985 | 6.457 | 24.552 | 35.392 | 18.267 | 8.711 | 4.201 | 2.264 | 1.063 | 0.217 | 0.233 | 0.102 | 0.038 |
| 1986 | 20.642 | 20.330 | 26.644 | 30.839 | 11.413 | 4.441 | 1.771 | 0.805 | 0.392 | 0.103 | 0.076 | 0.044 |
| 1987 | 11.002 | 62.130 | 27.192 | 15.127 | 15.695 | 4.159 | 1.463 | 0.592 | 0.253 | 0.142 | 0.046 | 0.058 |
| 1988 | 6.713 | 39.323 | 55.895 | 18.663 | 6.399 | 5.877 | 1.345 | 0.455 | 0.305 | 0.157 | 0.114 | 0.025 |
| 1989 | 2.605 | 27.983 | 50.059 | 31.455 | 6.010 | 1.915 | 0.881 | 0.225 | 0.107 | 0.086 | 0.038 | 0.005 |
| 1990 | 5.785 | 12.313 | 27.179 | 44.534 | 17.037 | 2.573 | 0.609 | 0.322 | 0.118 | 0.050 | 0.015 | 0.020 |
| 1991 | 8.554 | 25.131 | 15.491 | 21.514 | 25.038 | 6.364 | 0.903 | 0.243 | 0.125 | 0.063 | 0.011 | 0.012 |
| 1992 | 12.217 | 21.708 | 26.524 | 11.413 | 10.073 | 8.304 | 2.006 | 0.257 | 0.046 | 0.032 | 0.009 | 0.008 |
| 1993 | 20.500 | 33.078 | 15.195 | 13.281 | 3.583 | 2.785 | 2.707 | 1.181 | 0.180 | 0.034 | 0.011 | 0.013 |
| 1994 | 6.160 | 24.142 | 19.666 | 6.968 | 4.393 | 1.257 | 0.599 | 0.508 | 0.283 | 0.049 | 0.018 | 0.006 |
| 1995 | 10.770 | 9.103 | 16.829 | 13.066 | 4.115 | 1.596 | 0.313 | 0.184 | 0.156 | 0.141 | 0.029 | 0.008 |
| 1996 | 5.356 | 14.886 | 7.372 | 12.307 | 9.429 | 2.157 | 0.837 | 0.208 | 0.076 | 0.065 | 0.055 | 0.005 |
| 1997 | 1.722 | 16.442 | 17.298 | 6.711 | 7.379 | 5.958 | 1.147 | 0.493 | 0.126 | 0.028 | 0.037 | 0.021 |
| 1998 | 3.458 | 7.707 | 25.394 | 20.167 | 5.893 | 3.856 | 2.951 | 0.500 | 0.196 | 0.055 | 0.033 | 0.013 |
| 1999 | 2.525 | 19.554 | 15.226 | 24.622 | 12.966 | 2.795 | 1.489 | 0.748 | 0.140 | 0.046 | 0.010 | 0.005 |
| 2000 | 10.493 | 6.581 | 29.080 | 11.227 | 11.390 | 5.714 | 1.104 | 0.567 | 0.314 | 0.074 | 0.022 | 0.006 |
| 2001 | 13.566 | 25.935 | 9.107 | 20.183 | 5.848 | 3.763 | 2.029 | 0.508 | 0.199 | 0.136 | 0.013 | 0.031 |
| 2002 | 5.992 | 17.762 | 24.056 | 7.168 | 9.430 | 2.453 | 1.556 | 0.739 | 0.150 | 0.058 | 0.041 | 0.004 |
| 2003 | 5.489 | 16.312 | 22.045 | 16.629 | 4.840 | 4.933 | 1.201 | 0.507 | 0.211 | 0.046 | 0.026 | 0.033 |
| 2004 | 1.784 | 17.958 | 24.043 | 17.903 | 10.167 | 2.881 | 1.977 | 0.500 | 0.162 | 0.087 | 0.019 | 0.008 |


| YEAR | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 5.236 | 5.283 | 26.129 | 16.952 | 8.577 | 4.901 | 1.295 | 0.790 | 0.217 | 0.096 | 0.037 | 0.005 |
| 2006 | 3.456 | 13.066 | 8.784 | 21.926 | 10.577 | 4.703 | 2.170 | 0.472 | 0.241 | 0.040 | 0.016 | 0.010 |
| 2007 | 2.034 | 11.540 | 15.826 | 8.563 | 9.904 | 5.730 | 2.299 | 1.150 | 0.332 | 0.088 | 0.067 | 0.006 |
| 2008 | 3.109 | 5.118 | 12.808 | 11.597 | 5.141 | 4.700 | 2.138 | 0.881 | 0.279 | 0.069 | 0.044 | 0.004 |
| 2009 | 3.448 | 7.892 | 9.571 | 17.860 | 10.474 | 3.888 | 2.306 | 0.744 | 0.316 | 0.089 | 0.023 | 0.012 |
| 2010 | 3.498 | 7.673 | 9.478 | 8.407 | 10.953 | 5.561 | 1.567 | 0.927 | 0.297 | 0.145 | 0.063 | 0.017 |
| 2011 | 4.014 | 7.832 | 10.522 | 10.788 | 6.281 | 6.300 | 2.418 | 0.678 | 0.419 | 0.135 | 0.039 | 0.016 |
| 2012 | 4.072 | 11.276 | 10.795 | 9.494 | 8.896 | 5.011 | 3.202 | 1.148 | 0.291 | 0.225 | 0.079 | 0.026 |
| 2013 | 5.780 | 12.243 | 15.364 | 11.413 | 7.589 | 5.789 | 2.571 | 1.832 | 0.653 | 0.209 | 0.146 | 0.036 |
| 2014 | 4.623 | 8.378 | 14.913 | 13.288 | 8.427 | 4.928 | 2.814 | 1.393 | 0.964 | 0.376 | 0.127 | 0.104 |
| 2015 | 5.164 | 13.257 | 10.289 | 13.989 | 9.552 | 5.658 | 2.450 | 1.556 | 0.953 | 0.409 | 0.125 | 0.037 |

Table 2. Icelandic cod in Division 5.a. Estimated mean weight at age in the landings ( $\mathbf{k g}$ ) in period the 1955-2015. The weights for age groups 3-9 in 2016 are based on predictions from the 2016 spring survey measurements. The weights in the catches are used to calculate the reference biomass ( $\boldsymbol{B}_{4+}$ ).

| YEAR | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 0.827 | 1.307 | 2.157 | 3.617 | 4.638 | 5.657 | 6.635 | 6.168 | 8.746 | 8.829 | 10.086 | 14.584 |
| 1956 | 1.080 | 1.600 | 2.190 | 3.280 | 4.650 | 5.630 | 6.180 | 6.970 | 6.830 | 9.290 | 10.965 | 12.954 |
| 1957 | 1.140 | 1.710 | 2.520 | 3.200 | 4.560 | 5.960 | 7.170 | 7.260 | 8.300 | 8.290 | 10.350 | 13.174 |
| 1958 | 1.210 | 1.810 | 3.120 | 4.510 | 5.000 | 5.940 | 6.640 | 8.290 | 8.510 | 8.840 | 9.360 | 13.097 |
| 1959 | 1.110 | 1.950 | 2.930 | 4.520 | 5.520 | 6.170 | 6.610 | 7.130 | 8.510 | 8.670 | 9.980 | 11.276 |
| 1960 | 1.060 | 1.720 | 2.920 | 4.640 | 5.660 | 6.550 | 6.910 | 7.140 | 7.970 | 10.240 | 10.100 | 12.871 |
| 1961 | 1.020 | 1.670 | 2.700 | 4.330 | 5.530 | 6.310 | 6.930 | 7.310 | 7.500 | 8.510 | 9.840 | 14.550 |
| 1962 | 0.990 | 1.610 | 2.610 | 3.900 | 5.720 | 6.660 | 6.750 | 7.060 | 7.540 | 8.280 | 10.900 | 12.826 |
| 1963 | 1.250 | 1.650 | 2.640 | 3.800 | 5.110 | 6.920 | 7.840 | 7.610 | 8.230 | 9.100 | 9.920 | 11.553 |
| 1964 | 1.210 | 1.750 | 2.640 | 4.020 | 5.450 | 6.460 | 8.000 | 9.940 | 9.210 | 10.940 | 12.670 | 15.900 |
| 1965 | 1.020 | 1.530 | 2.570 | 4.090 | 5.410 | 6.400 | 7.120 | 8.600 | 12.310 | 10.460 | 10.190 | 17.220 |
| 1966 | 1.170 | 1.680 | 2.590 | 4.180 | 5.730 | 6.900 | 7.830 | 8.580 | 9.090 | 14.230 | 14.090 | 17.924 |
| 1967 | 1.120 | 1.820 | 2.660 | 4.067 | 5.560 | 7.790 | 7.840 | 8.430 | 9.090 | 10.090 | 14.240 | 16.412 |
| 1968 | 1.170 | 1.590 | 2.680 | 3.930 | 5.040 | 5.910 | 7.510 | 8.480 | 10.750 | 11.580 | 14.640 | 16.011 |
| 1969 | 1.100 | 1.810 | 2.480 | 3.770 | 5.040 | 5.860 | 7.000 | 8.350 | 8.720 | 10.080 | 11.430 | 13.144 |
| 1970 | 0.990 | 1.450 | 2.440 | 3.770 | 4.860 | 5.590 | 6.260 | 8.370 | 10.490 | 12.310 | 14.590 | 21.777 |
| 1971 | 1.090 | 1.570 | 2.310 | 2.980 | 4.930 | 5.150 | 5.580 | 6.300 | 8.530 | 11.240 | 14.740 | 17.130 |
| 1972 | 0.980 | 1.460 | 2.210 | 3.250 | 4.330 | 5.610 | 6.040 | 6.100 | 6.870 | 8.950 | 11.720 | 16.000 |
| 1973 | 1.030 | 1.420 | 2.470 | 3.600 | 4.900 | 6.110 | 6.670 | 6.750 | 7.430 | 7.950 | 10.170 | 17.000 |
| 1974 | 1.050 | 1.710 | 2.430 | 3.820 | 5.240 | 6.660 | 7.150 | 7.760 | 8.190 | 9.780 | 12.380 | 14.700 |
| 1975 | 1.100 | 1.770 | 2.780 | 3.760 | 5.450 | 6.690 | 7.570 | 8.580 | 8.810 | 9.780 | 10.090 | 11.000 |
| 1976 | 1.350 | 1.780 | 2.650 | 4.100 | 5.070 | 6.730 | 8.250 | 9.610 | 11.540 | 11.430 | 14.060 | 16.180 |
| 1977 | 1.259 | 1.911 | 2.856 | 4.069 | 5.777 | 6.636 | 7.685 | 9.730 | 11.703 | 14.394 | 17.456 | 24.116 |
| 1978 | 1.289 | 1.833 | 2.929 | 3.955 | 5.726 | 6.806 | 9.041 | 10.865 | 13.068 | 11.982 | 19.062 | 21.284 |
| 1979 | 1.408 | 1.956 | 2.642 | 3.999 | 5.548 | 6.754 | 8.299 | 9.312 | 13.130 | 13.418 | 13.540 | 20.072 |
| 1980 | 1.392 | 1.862 | 2.733 | 3.768 | 5.259 | 6.981 | 8.037 | 10.731 | 12.301 | 17.281 | 14.893 | 19.069 |
| 1981 | 1.180 | 1.651 | 2.260 | 3.293 | 4.483 | 5.821 | 7.739 | 9.422 | 11.374 | 12.784 | 12.514 | 19.069 |
| 1982 | 1.006 | 1.550 | 2.246 | 3.104 | 4.258 | 5.386 | 6.682 | 9.141 | 11.963 | 14.226 | 17.287 | 16.590 |
| 1983 | 1.095 | 1.599 | 2.275 | 3.021 | 4.096 | 5.481 | 7.049 | 8.128 | 11.009 | 13.972 | 15.882 | 18.498 |
| 1984 | 1.288 | 1.725 | 2.596 | 3.581 | 4.371 | 5.798 | 7.456 | 9.851 | 11.052 | 14.338 | 15.273 | 16.660 |
| 1985 | 1.407 | 1.971 | 2.576 | 3.650 | 4.976 | 6.372 | 8.207 | 10.320 | 12.197 | 14.683 | 16.175 | 19.050 |
| 1986 | 1.459 | 1.961 | 2.844 | 3.593 | 4.635 | 6.155 | 7.503 | 9.084 | 10.356 | 15.283 | 14.540 | 15.017 |
| 1987 | 1.316 | 1.956 | 2.686 | 3.894 | 4.716 | 6.257 | 7.368 | 9.243 | 10.697 | 10.622 | 15.894 | 12.592 |
| 1988 | 1.438 | 1.805 | 2.576 | 3.519 | 4.930 | 6.001 | 7.144 | 8.822 | 9.977 | 11.732 | 14.156 | 13.042 |
| 1989 | 1.186 | 1.813 | 2.590 | 3.915 | 5.210 | 6.892 | 8.035 | 9.831 | 11.986 | 10.003 | 12.611 | 16.045 |
| 1990 | 1.290 | 1.704 | 2.383 | 3.034 | 4.624 | 6.521 | 8.888 | 10.592 | 10.993 | 14.570 | 15.732 | 17.290 |
| 1991 | 1.309 | 1.899 | 2.475 | 3.159 | 3.792 | 5.680 | 7.242 | 9.804 | 9.754 | 14.344 | 14.172 | 20.200 |
| 1992 | 1.289 | 1.768 | 2.469 | 3.292 | 4.394 | 5.582 | 6.830 | 8.127 | 12.679 | 13.410 | 15.715 | 11.267 |
| 1993 | 1.392 | 1.887 | 2.772 | 3.762 | 4.930 | 6.054 | 7.450 | 8.641 | 10.901 | 12.517 | 14.742 | 16.874 |
| 1994 | 1.443 | 2.063 | 2.562 | 3.659 | 5.117 | 6.262 | 7.719 | 8.896 | 10.847 | 12.874 | 14.742 | 17.470 |
| 1995 | 1.348 | 1.959 | 2.920 | 3.625 | 5.176 | 6.416 | 7.916 | 10.273 | 11.022 | 11.407 | 13.098 | 15.182 |
| 1996 | 1.457 | 1.930 | 3.132 | 4.141 | 4.922 | 6.009 | 7.406 | 9.772 | 10.539 | 13.503 | 13.689 | 16.194 |
| 1997 | 1.484 | 1.877 | 2.878 | 4.028 | 5.402 | 6.386 | 7.344 | 8.537 | 10.797 | 11.533 | 10.428 | 12.788 |
| 1998 | 1.230 | 1.750 | 2.458 | 3.559 | 5.213 | 7.737 | 7.837 | 9.304 | 10.759 | 14.903 | 16.651 | 18.666 |
| 1999 | 1.241 | 1.716 | 2.426 | 3.443 | 4.720 | 6.352 | 8.730 | 9.946 | 11.088 | 12.535 | 14.995 | 15.151 |
| 2000 | 1.308 | 1.782 | 2.330 | 3.252 | 4.690 | 5.894 | 7.809 | 9.203 | 10.240 | 11.172 | 13.172 | 17.442 |
| 2001 | 1.486 | 2.021 | 2.631 | 3.364 | 4.558 | 6.190 | 7.124 | 8.445 | 9.313 | 9.569 | 10.234 | 9.505 |
| 2002 | 1.308 | 1.946 | 2.662 | 3.636 | 4.550 | 5.927 | 7.082 | 8.100 | 9.275 | 11.660 | 11.220 | 14.025 |
| 2003 | 1.350 | 1.866 | 2.459 | 3.391 | 4.380 | 4.756 | 6.141 | 7.138 | 9.580 | 10.260 | 11.479 | 10.720 |
| 2004 | 1.139 | 1.754 | 2.413 | 3.372 | 4.288 | 5.185 | 5.740 | 7.376 | 10.037 | 10.322 | 12.428 | 11.445 |
| 2005 | 1.195 | 1.734 | 2.419 | 3.392 | 4.292 | 5.057 | 6.232 | 6.123 | 7.961 | 10.067 | 12.776 | 13.717 |


| YEAR | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 1.089 | 1.625 | 2.210 | 3.059 | 4.270 | 4.983 | 5.290 | 6.040 | 8.448 | 11.155 | 12.611 | 15.382 |
| 2007 | 1.062 | 1.593 | 2.179 | 2.791 | 3.865 | 5.162 | 5.876 | 6.407 | 7.186 | 9.519 | 10.408 | 10.532 |
| 2008 | 1.100 | 1.600 | 2.369 | 3.147 | 3.996 | 5.278 | 6.495 | 7.383 | 7.822 | 10.391 | 11.562 | 18.087 |
| 2009 | 1.096 | 1.668 | 2.210 | 3.190 | 4.068 | 5.035 | 6.663 | 8.371 | 9.520 | 11.205 | 11.753 | 15.036 |
| 2010 | 1.100 | 1.827 | 2.360 | 3.222 | 4.485 | 5.471 | 6.748 | 8.038 | 8.975 | 10.395 | 11.629 | 12.222 |
| 2011 | 1.111 | 1.664 | 2.517 | 3.452 | 4.412 | 5.792 | 6.531 | 7.826 | 8.810 | 9.697 | 12.942 | 11.644 |
| 2012 | 1.184 | 1.631 | 2.452 | 3.760 | 4.717 | 5.934 | 7.368 | 8.011 | 9.098 | 10.718 | 12.037 | 11.596 |
| 2013 | 1.132 | 1.743 | 2.450 | 3.611 | 4.936 | 6.126 | 7.368 | 8.137 | 9.173 | 10.121 | 10.422 | 12.703 |
| 2014 | 1.117 | 1.740 | 2.521 | 3.515 | 4.675 | 6.158 | 7.486 | 8.583 | 8.962 | 10.516 | 10.281 | 12.324 |
| 2015 | 1.188 | 1.629 | 2.635 | 3.584 | 4.633 | 5.909 | 7.573 | 8.603 | 9.688 | 11.212 | 11.334 | 10.356 |
| 2016 | 1.401 | 1.815 | 2.265 | 3.500 | 4.153 | 5.338 | 6.664 | 8.603 | 9.688 | 11.212 | 11.334 | 10.356 |

Table 3. Icelandic cod in Division 5.a. Estimated weight at age in the spawning stock ( $\mathbf{k g}$ ) in period the 1955-2017. These weights are used to calculate the spawning stock biomass (SSB).

| YEAR | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 0.645 | 1.019 | 1.833 | 3.183 | 4.128 | 5.657 | 6.635 | 6.168 | 8.746 | 8.829 | 10.086 | 14.584 |
| 1956 | 0.645 | 1.248 | 1.862 | 2.886 | 4.138 | 5.630 | 6.180 | 6.970 | 6.830 | 9.290 | 10.965 | 12.954 |
| 1957 | 0.645 | 1.334 | 2.142 | 2.816 | 4.058 | 5.960 | 7.170 | 7.260 | 8.300 | 8.290 | 10.350 | 13.174 |
| 1958 | 0.645 | 1.412 | 2.652 | 3.969 | 4.450 | 5.940 | 6.640 | 8.290 | 8.510 | 8.840 | 9.360 | 13.097 |
| 1959 | 0.645 | 1.521 | 2.490 | 3.978 | 4.913 | 6.170 | 6.610 | 7.130 | 8.510 | 8.670 | 9.980 | 11.276 |
| 1960 | 0.645 | 1.342 | 2.482 | 4.083 | 5.037 | 6.550 | 6.910 | 7.140 | 7.970 | 10.240 | 10.100 | 12.871 |
| 1961 | 0.645 | 1.303 | 2.295 | 3.810 | 4.922 | 6.310 | 6.930 | 7.310 | 0.750 | 8.510 | 9.840 | 14.550 |
| 1962 | 0.645 | 1.256 | 2.218 | 3.432 | 5.091 | 6.660 | 6.750 | 7.060 | 7.540 | 8.280 | 10.900 | 12.826 |
| 1963 | 0.645 | 1.287 | 2.244 | 3.344 | 4.548 | 6.920 | 7.840 | 7.610 | 8.230 | 9.100 | 9.920 | 11.553 |
| 1964 | 0.645 | 1.365 | 2.244 | 3.538 | 4.850 | 6.460 | 8.000 | 9.940 | 9.210 | 10.940 | 12.670 | 15.900 |
| 1965 | 0.645 | 1.193 | 2.184 | 3.59 | 4.815 | 6.400 | 7.120 | 8.600 | 12.310 | 10.460 | 10.190 | 17.220 |
| 1966 | 0.645 | 1.310 | 2.202 | 3.67 | 5.10 | 6.900 | 7.830 | 8.580 | 9.090 | 14.230 | 14.090 | 17.924 |
| 1967 | 0.645 | 1.420 | 2.261 | 3.579 | 4.948 | 7.790 | 7.840 | 8.430 | 9.090 | 10.090 | 14.240 | 16.412 |
| 1968 | 0.645 | 1.240 | 2.278 | 3.458 | 4.486 | 5.910 | 7.510 | 8.480 | 10.750 | 11.580 | 14.640 | 16.011 |
| 1969 | 0.645 | 1.412 | 2.108 | 3.318 | 4.486 | 5.860 | 7.000 | 8.350 | 8.720 | 10.080 | 11.430 | 13.144 |
| 1970 | 0.645 | 1.131 | 2.074 | 3.318 | 4.325 | 5.590 | 6.260 | 8.370 | 10.490 | 12.310 | 14.590 | 21.777 |
| 1971 | 0.645 | 1.225 | 1.964 | 2.62 | 4.38 | 5.15 | 5.580 | 6.300 | 8.530 | 11.240 | 14.740 | 17.130 |
| 19 | 0.645 | 1.139 | 1.878 | 2.86 | 3.85 | 5.61 | 6.040 | 6.100 | 6.870 | 8.950 | 11.720 | 16.000 |
| 1973 | 0.64 | 1.108 | 2.100 | 3.16 | 4.36 | 6.11 | 6.670 | 6.750 | 7.430 | 7.950 | 10.170 | 17.000 |
| 1974 | 0.645 | 1.334 | 2.066 | 3.36 | 4.66 | 6.660 | 7.150 | 7.760 | 8.190 | 9.780 | 12.380 | 14.700 |
| 1975 | 0.645 | 1.381 | 2.363 | 3.30 | 4.850 | 6.690 | 7.570 | 8.580 | 8.810 | 9.780 | 10.090 | 11.000 |
| 1976 | 0.645 | 1.388 | 2.252 | 3.608 | 4.512 | 6.730 | 8.250 | 9.610 | 11.540 | 11.430 | 14.060 | 16.180 |
| 1977 | 0.645 | 1.491 | 2.428 | 3.581 | 5.142 | 6.636 | 7.685 | 9.730 | 11.703 | 14.394 | 17.456 | 24.116 |
| 1978 | 0.645 | 1.430 | 2.490 | 3.480 | 5.096 | 6.806 | 9.041 | 10.865 | 13.068 | 11.982 | 19.062 | 21.284 |
| 1979 | 0.645 | 1.526 | 2.246 | 3.519 | 4.938 | 6.754 | 8.299 | 9.312 | 13.130 | 13.418 | 13.540 | 20.072 |
| 1980 | 0.645 | 1.452 | 2.323 | 3.31 | 4.681 | 6.981 | 8.037 | 10.731 | 12.301 | 17.281 | 14.893 | 19.069 |
| 1981 | 0.645 | 1.288 | 1.921 | 2.898 | 3.990 | 5.821 | 7.739 | 9.422 | 11.374 | 12.784 | 12.514 | 19.069 |
| 1982 | 0.645 | 1.209 | 1.909 | 2.732 | 3.790 | 5.386 | 6.682 | 9.141 | 11.963 | 14.226 | 17.287 | 16.590 |
| 1983 | 0.645 | 1.247 | 1.934 | 2.658 | 3.645 | 5.481 | 7.049 | 8.128 | 11.009 | 13.972 | 15.882 | 18.498 |
| 1984 | 0.645 | 1.346 | 2.207 | 3.151 | 3.890 | 5.798 | 7.456 | 9.851 | 11.052 | 14.338 | 15.273 | 16.660 |
| 1985 | 1.306 | 1.382 | 1.752 | 2.710 | 3.443 | 4.675 | 7.220 | 10.320 | 12.197 | 14.683 | 16.175 | 19.050 |
| 1986 | 1.306 | 1.604 | 2.892 | 3.234 | 4.572 | 5.805 | 7.247 | 9.084 | 10.356 | 15.283 | 14.540 | 15.017 |
| 1987 | 1.706 | 1.589 | 2.426 | 3.516 | 4.879 | 6.459 | 7.656 | 9.243 | 10.697 | 10.622 | 15.894 | 12.592 |
| 1988 | 0.929 | 1.480 | 2.263 | 3.273 | 4.387 | 4.566 | 8.275 | 8.822 | 9.977 | 11.732 | 14.156 | 13.042 |
| 1989 | 0.822 | 1.501 | 2.346 | 3.428 | 4.676 | 7.388 | 8.506 | 9.831 | 11.986 | 10.003 | 12.611 | 16.045 |
| 1990 | 0.725 | 1.043 | 2.179 | 2.809 | 4.421 | 6.359 | 9.230 | 10.592 | 10.993 | 14.570 | 15.732 | 17.290 |
| 1991 | 0.114 | 1.286 | 2.042 | 2.752 | 3.404 | 6.091 | 9.152 | 9.804 | 9.754 | 14.344 | 14.172 | 20.200 |
| 1992 | 0.448 | 1.344 | 2.096 | 3.029 | 3.755 | 5.143 | 7.562 | 8.127 | 12.679 | 13.410 | 15.715 | 11.267 |
| 1993 | 0.773 | 1.363 | 2.309 | 3.236 | 4.111 | 5.710 | 6.352 | 8.641 | 10.901 | 12.517 | 14.742 | 16.874 |
| 1994 | 1.611 | 1.728 | 2.253 | 3.341 | 4.515 | 6.535 | 10.039 | 8.896 | 10.847 | 12.874 | 14.742 | 17.470 |
| 1995 | 0.514 | 1.636 | 2.346 | 3.186 | 4.488 | 5.528 | 8.620 | 10.273 | 11.022 | 11.407 | 13.098 | 15.182 |
| 1996 | 0.543 | 1.754 | 2.491 | 3.534 | 4.254 | 5.634 | 8.300 | 9.772 | 10.539 | 13.503 | 13.689 | 16.193 |


| YEAR | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 1.112 | 1.347 | 2.267 | 3.746 | 5.426 | 5.972 | 6.958 | 8.537 | 10.797 | 11.533 | 10.428 | 12.788 |
| 1998 | 1.112 | 1.821 | 2.261 | 3.263 | 4.468 | 5.784 | 6.812 | 9.304 | 10.759 | 14.903 | 16.651 | 18.666 |
| 1999 | 1.307 | 1.467 | 1.933 | 2.997 | 3.961 | 5.120 | 6.494 | 9.946 | 11.088 | 12.535 | 14.995 | 15.151 |
| 2000 | 0.496 | 1.355 | 1.916 | 2.881 | 4.318 | 5.580 | 8.497 | 9.203 | 10.240 | 11.172 | 13.172 | 17.442 |
| 2001 | 0.816 | 1.583 | 2.108 | 2.700 | 4.086 | 6.202 | 6.907 | 9.055 | 8.769 | 9.526 | 11.210 | 13.874 |
| 2002 | 0.780 | 1.590 | 2.259 | 3.120 | 3.985 | 5.958 | 9.234 | 9.002 | 10.422 | 13.402 | 9.008 | 16.893 |
| 2003 | 1.149 | 1.324 | 2.239 | 3.052 | 4.231 | 5.057 | 6.838 | 7.819 | 8.802 | 10.712 | 12.152 | 13.797 |
| 2004 | 1.149 | 1.430 | 2.099 | 3.049 | 3.743 | 5.319 | 5.682 | 7.397 | 10.808 | 11.569 | 13.767 | 12.955 |
| 2005 | 0.649 | 1.120 | 1.898 | 2.962 | 3.875 | 4.806 | 7.281 | 5.495 | 7.211 | 9.909 | 12.944 | 18.151 |
| 2006 | 0.907 | 1.384 | 1.999 | 2.907 | 4.384 | 5.122 | 6.536 | 5.769 | 6.258 | 5.688 | 7.301 | 15.412 |
| 2007 | 1.403 | 1.264 | 2.022 | 2.582 | 4.081 | 5.725 | 6.736 | 6.481 | 7.142 | 6.530 | 9.724 | 10.143 |
| 2008 | 0.912 | 1.842 | 2.232 | 2.925 | 3.915 | 5.462 | 7.075 | 7.648 | 8.282 | 11.181 | 14.266 | 17.320 |
| 2009 | 0.644 | 1.441 | 2.028 | 2.873 | 3.913 | 4.919 | 7.046 | 8.505 | 10.126 | 12.108 | 12.471 | 15.264 |
| 2010 | 0.644 | 1.588 | 2.153 | 3.131 | 4.173 | 5.197 | 6.356 | 7.945 | 8.913 | 10.090 | 10.417 | 13.489 |
| 2011 | 0.794 | 2.377 | 2.651 | 3.203 | 4.517 | 6.000 | 6.866 | 7.850 | 8.810 | 9.797 | 13.534 | 13.033 |
| 2012 | 1.403 | 1.698 | 2.594 | 3.683 | 4.483 | 5.921 | 7.988 | 8.358 | 9.543 | 10.916 | 10.884 | 11.758 |
| 2013 | 0.944 | 2.282 | 2.983 | 3.827 | 5.206 | 6.543 | 8.298 | 8.415 | 9.336 | 9.926 | 11.195 | 12.691 |
| 2014 | 0.944 | 1.333 | 2.539 | 3.307 | 4.460 | 6.424 | 8.225 | 8.413 | 9.713 | 10.513 | 11.437 | 12.979 |
| 2015 | 0.709 | 1.046 | 3.308 | 3.829 | 4.897 | 6.234 | 8.719 | 9.694 | 9.688 | 11.212 | 11.334 | 10.356 |
| 2016 | 0.973 | 2.223 | 3.035 | 4.198 | 4.610 | 6.018 | 7.389 | 9.731 | 9.688 | 11.212 | 11.334 | 10.356 |

Table 4. Icelandic cod in Division 5.a. Estimated maturity at age in period the 1955-2016.

| YEAR | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 0.019 | 0.022 | 0.033 | 0.181 | 0.577 | 0.782 | 0.834 | 0.960 | 1.000 | 1.000 | 1.000 |
| 1956 | 0.019 | 0.025 | 0.033 | 0.111 | 0.577 | 0.782 | 0.818 | 0.980 | 0.980 | 1.000 | 1.000 |
| 1957 | 0.019 | 0.026 | 0.043 | 0.100 | 0.549 | 0.801 | 0.842 | 0.990 | 1.000 | 1.000 | 1.000 |
| 1958 | 0.019 | 0.028 | 0.086 | 0.520 | 0.682 | 0.801 | 0.834 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1959 | 0.019 | 0.029 | 0.070 | 0.535 | 0.772 | 0.818 | 0.834 | 0.990 | 1.000 | 1.000 | 1.000 |
| 1960 | 0.019 | 0.026 | 0.066 | 0.577 | 0.782 | 0.826 | 0.834 | 0.990 | 1.000 | 1.000 | 1.000 |
| 1961 | 0.019 | 0.025 | 0.053 | 0.450 | 0.772 | 0.818 | 0.834 | 0.990 | 0.990 | 1.000 | 1.000 |
| 1962 | 0.019 | 0.025 | 0.048 | 0.281 | 0.79 | 0.834 | 0.834 | 0.990 | 0.990 | 1.000 | 1.000 |
| 1963 | 0.019 | 0.025 | 0.048 | 0.237 | 0.706 | 0.834 | 0.849 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1964 | 0.019 | 0.026 | 0.048 | 0.329 | 0.76 | 0.826 | 0.849 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1965 | 0.019 | 0.025 | 0.045 | 0.354 | 0.751 | 0.826 | 0.842 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1966 | 0.019 | 0.026 | 0.045 | 0.394 | 0.791 | 0.849 | 0.849 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1967 | 0.019 | 0.028 | 0.051 | 0.341 | 0.772 | 0.842 | 0.849 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1968 | 0.019 | 0.025 | 0.051 | 0.292 | 0.682 | 0.801 | 0.842 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1969 | 0.019 | 0.028 | 0.043 | 0.227 | 0.682 | 0.801 | 0.842 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1970 | 0.019 | 0.023 | 0.04 | 0.227 | 0.644 | 0.772 | 0.818 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1971 | 0.019 | 0.025 | 0.03 | 0.074 | 0.65 | 0.706 | 0.772 | 0.979 | 0.994 | 0.982 | 0.993 |
| 1972 | 0.019 | 0.023 | 0.03 | 0.106 | 0.45 | 0.772 | 0.809 | 0.979 | 0.994 | 0.982 | 0.993 |
| 1973 | 0.022 | 0.028 | 0.16 | 0.382 | 0.697 | 0.801 | 0.834 | 0.996 | 0.996 | 1.000 | 1.000 |
| 1974 | 0.020 | 0.031 | 0.085 | 0.346 | 0.636 | 0.790 | 0.818 | 0.989 | 1.000 | 1.000 | 1.000 |
| 1975 | 0.020 | 0.035 | 0.118 | 0.287 | 0.715 | 0.809 | 0.839 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1976 | 0.025 | 0.026 | 0.086 | 0.253 | 0.406 | 0.797 | 0.841 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1977 | 0.019 | 0.024 | 0.060 | 0.382 | 0.742 | 0.817 | 0.842 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1978 | 0.025 | 0.025 | 0.05 | 0.192 | 0.737 | 0.820 | 0.836 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1979 | 0.019 | 0.021 | 0.05 | 0.282 | 0.635 | 0.790 | 0.836 | 0.919 | 1.000 | 1.000 | 1.000 |
| 1980 | 0.026 | 0.021 | 0.04 | 0.225 | 0.653 | 0.777 | 0.834 | 0.977 | 1.000 | 0.964 | 1.000 |
| 1981 | 0.019 | 0.022 | 0.030 | 0.090 | 0.448 | 0.751 | 0.811 | 0.962 | 0.988 | 1.000 | 1.000 |
| 1982 | 0.021 | 0.025 | 0.038 | 0.065 | 0.297 | 0.705 | 0.815 | 0.967 | 1.000 | 1.000 | 1.000 |
| 1983 | 0.019 | 0.030 | 0.047 | 0.116 | 0.264 | 0.530 | 0.715 | 0.979 | 0.985 | 1.000 | 1.000 |
| 1984 | 0.019 | 0.024 | 0.053 | 0.169 | 0.444 | 0.620 | 0.716 | 0.949 | 0.969 | 0.948 | 1.000 |
| 1985 | NA | 0.021 | 0.185 | 0.412 | 0.495 | 0.735 | 0.572 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1986 | 0.001 | 0.023 | 0.149 | 0.395 | 0.682 | 0.734 | 0.941 | 0.962 | 0.988 | 1.000 | 1.000 |
| 1987 | 0.002 | 0.033 | 0.093 | 0.360 | 0.490 | 0.885 | 0.782 | 1.000 | 0.979 | 1.000 | 1.000 |
| 1988 | 0.006 | 0.029 | 0.225 | 0.511 | 0.448 | 0.683 | 0.937 | 0.946 | 0.974 | 0.821 | 1.000 |
| 1989 | 0.008 | 0.025 | 0.142 | 0.372 | 0.645 | 0.652 | 0.634 | 0.991 | 1.000 | 0.903 | 0.859 |
| 1990 | 0.006 | 0.012 | 0.155 | 0.437 | 0.581 | 0.796 | 0.814 | 0.986 | 1.000 | 1.000 | 1.000 |
| 1991 | NA | 0.055 | 0.149 | 0.369 | 0.637 | 0.790 | 0.682 | 0.842 | 1.000 | 1.000 | 1.000 |
| 1992 | 0.002 | 0.062 | 0.265 | 0.402 | 0.813 | 0.917 | 0.894 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1993 | 0.006 | 0.085 | 0.267 | 0.464 | 0.693 | 0.801 | 0.843 | 0.968 | 1.000 | 1.000 | 1.000 |
| 1994 | 0.008 | 0.110 | 0.339 | 0.591 | 0.702 | 0.917 | 0.698 | 0.852 | 0.985 | 1.000 | 1.000 |
| 1995 | 0.005 | 0.109 | 0.384 | 0.528 | 0.752 | 0.787 | 0.859 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1996 | 0.002 | 0.031 | 0.186 | 0.499 | 0.650 | 0.733 | 0.812 | 1.000 | 1.000 | 0.986 | 0.971 |
| 1997 | 0.006 | 0.037 | 0.246 | 0.424 | 0.685 | 0.787 | 0.804 | 0.932 | 1.000 | 0.913 | 1.000 |


| YEAR | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | NA | 0.061 | 0.209 | 0.491 | 0.782 | 0.814 | 0.810 | 0.925 | 0.998 | 1.000 | 1.000 |
| 1999 | 0.012 | 0.044 | 0.239 | 0.516 | 0.649 | 0.835 | 0.687 | 0.988 | 1.000 | 1.000 | 1.000 |
| 2000 | 0.001 | 0.065 | 0.248 | 0.512 | 0.611 | 0.867 | 0.998 | 0.980 | 1.000 | 1.000 | 1.000 |
| 2001 | 0.004 | 0.043 | 0.261 | 0.589 | 0.750 | 0.742 | 0.862 | 0.987 | 1.000 | 1.000 | 1.000 |
| 2002 | 0.008 | 0.086 | 0.322 | 0.656 | 0.759 | 0.920 | 0.550 | 0.979 | 1.000 | 1.000 | 1.000 |
| 2003 | 0.005 | 0.046 | 0.218 | 0.524 | 0.870 | 0.798 | 0.860 | 0.998 | 1.000 | 1.000 | 1.000 |
| 2004 | NA | 0.038 | 0.246 | 0.549 | 0.626 | 0.843 | 0.816 | 0.990 | 1.000 | 1.000 | 1.000 |
| 2005 | 0.003 | 0.109 | 0.281 | 0.493 | 0.792 | 0.805 | 0.951 | 0.908 | 1.000 | 1.000 | 1.000 |
| 2006 | 0.002 | 0.023 | 0.294 | 0.448 | 0.752 | 0.871 | 0.743 | 0.747 | 1.000 | 1.000 | 1.000 |
| 2007 | 0.012 | 0.032 | 0.159 | 0.501 | 0.693 | 0.785 | 0.836 | 0.924 | 1.000 | 1.000 | 1.000 |
| 2008 | 0.001 | 0.041 | 0.276 | 0.549 | 0.727 | 0.827 | 0.846 | 0.954 | 1.000 | 1.000 | 1.000 |
| 2009 | 0.002 | 0.015 | 0.132 | 0.456 | 0.688 | 0.883 | 0.741 | 0.631 | 1.000 | 1.000 | 1.000 |
| 2010 | NA | 0.016 | 0.058 | 0.377 | 0.822 | 0.869 | 0.923 | 0.802 | 1.000 | 1.000 | 1.000 |
| 2011 | 0.002 | 0.012 | 0.135 | 0.431 | 0.734 | 0.926 | 0.940 | 0.958 | 1.000 | 1.000 | 1.000 |
| 2012 | 0.004 | 0.029 | 0.126 | 0.411 | 0.728 | 0.882 | 0.961 | 0.830 | 1.000 | 1.000 | 1.000 |
| 2013 | 0.003 | 0.008 | 0.061 | 0.343 | 0.738 | 0.923 | 0.957 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2014 | NA | 0.026 | 0.068 | 0.236 | 0.614 | 0.893 | 0.967 | 0.957 | 1.000 | 1.000 | 1.000 |
| 2015 | 0.004 | 0.007 | 0.109 | 0.353 | 0.638 | 0.910 | 0.979 | 0.988 | 1.000 | 1.000 | 1.000 |
| 2016 | 0.001 | 0.009 | 0.024 | 0.288 | 0.543 | 0.733 | 0.942 | 0.987 | 1.000 | 1.000 | 1.000 |

Table 5. Icelandic cod in Division 5.a. Estimated survey weight at age (g) in the spring survey (SMB).

| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 14 | 137 | 388 | 1117 | 1733 | 2578 | 3221 | 4667 | 5858 | 7022 |
| 1986 | 15 | 159 | 616 | 1219 | 2246 | 2961 | 4326 | 5582 | 7216 | 8298 |
| 1987 | 14 | 117 | 467 | 1198 | 1751 | 2980 | 4194 | 6329 | 6932 | 10026 |
| 1988 | 11 | 122 | 495 | 1076 | 1964 | 3094 | 3549 | 4357 | 8119 | 9424 |
| 1989 | 22 | 150 | 548 | 1141 | 1932 | 3048 | 4384 | 6255 | 6999 | 12508 |
| 1990 | 19 | 135 | 460 | 1039 | 1815 | 2595 | 3871 | 6039 | 8141 | 9560 |
| 1991 | 18 | 147 | 553 | 1166 | 1842 | 2586 | 3267 | 5729 | 7601 | 14399 |
| 1992 | 24 | 134 | 501 | 1013 | 1845 | 2567 | 3651 | 5046 | 7433 | 13505 |
| 1993 | 12 | 171 | 576 | 1166 | 1944 | 2991 | 3961 | 5377 | 5983 | 9330 |
| 1994 | 13 | 174 | 686 | 1412 | 2044 | 3180 | 4132 | 6277 | 8310 | 10535 |
| 1995 | 10 | 134 | 605 | 1377 | 2284 | 2989 | 4449 | 5322 | 8063 | 9243 |
| 1996 | 11 | 155 | 551 | 1350 | 2082 | 3321 | 4044 | 5263 | 7475 | 9961 |
| 1997 | 18 | 140 | 546 | 1194 | 2168 | 3220 | 4863 | 5507 | 6457 | 6899 |
| 1998 | 15 | 158 | 485 | 1208 | 2041 | 3017 | 4253 | 5436 | 6346 | 8380 |
| 1999 | 14 | 140 | 578 | 1070 | 1847 | 2867 | 3819 | 4980 | 5625 | 8190 |
| 2000 | 16 | 124 | 486 | 1195 | 1817 | 2770 | 4066 | 5348 | 8499 | 8396 |
| 2001 | 17 | 152 | 531 | 1186 | 1852 | 2641 | 3760 | 5450 | 6440 | 8169 |
| 2002 | 11 | 132 | 510 | 1206 | 1998 | 2920 | 3779 | 5758 | 6259 | 6282 |
| 2003 | 16 | 131 | 466 | 1179 | 1918 | 2787 | 4139 | 4677 | 6258 | 9588 |
| 2004 | 20 | 147 | 481 | 1062 | 1873 | 2803 | 3458 | 4988 | 5312 | 7790 |
| 2005 | 11 | 118 | 451 | 1029 | 1760 | 2643 | 3646 | 4361 | 7246 | 6668 |
| 2006 | 13 | 105 | 417 | 982 | 1689 | 2600 | 4050 | 4749 | 5621 | 8374 |
| 2007 | 14 | 101 | 410 | 969 | 1663 | 2342 | 3635 | 5017 | 6120 | 7744 |
| 2008 | 11 | 121 | 376 | 937 | 1805 | 2612 | 3592 | 4932 | 6394 | 8404 |
| 2009 | 12 | 113 | 413 | 845 | 1602 | 2633 | 3659 | 4683 | 5768 | 6287 |
| 2010 | 13 | 98 | 391 | 1008 | 1697 | 2570 | 4021 | 4912 | 6100 | 7751 |
| 2011 | 12 | 102 | 395 | 1126 | 2114 | 2986 | 4225 | 5876 | 6644 | 7903 |
| 2012 | 12 | 142 | 477 | 1143 | 1929 | 3180 | 4249 | 5718 | 7825 | 7609 |
| 2013 | 13 | 111 | 495 | 1054 | 1785 | 3023 | 4774 | 6384 | 8114 | 9543 |
| 2014 | 11 | 114 | 359 | 1079 | 1710 | 2632 | 3987 | 6168 | 8068 | 10116 |
| 2015 | 13 | 150 | 418 | 898 | 2055 | 3016 | 4401 | 6074 | 8652 | 9618 |
| 2016 | 10 | 121 | 483 | 1013 | 1581 | 3146 | 3988 | 5522 | 7229 | 9683 |

Table 6. Icelandic cod in Division 5.a. Survey indices of the spring bottom trawl survey (SMB).

| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 16.54 | 110.43 | 35.40 | 48.20 | 64.15 | 22.57 | 14.85 | 4.85 | 3.21 | 1.76 |
| 1986 | 15.05 | 60.24 | 95.89 | 22.42 | 21.21 | 26.34 | 6.63 | 2.48 | 0.83 | 0.73 |
| 1987 | 3.65 | 28.21 | 103.74 | 81.99 | 21.08 | 12.21 | 12.00 | 2.55 | 0.89 | 0.38 |
| 1988 | 3.44 | 6.96 | 72.09 | 101.49 | 66.56 | 7.81 | 5.90 | 6.29 | 0.58 | 0.24 |
| 1989 | 4.04 | 16.38 | 21.97 | 77.79 | 67.59 | 34.20 | 4.20 | 1.45 | 1.14 | 0.24 |
| 1990 | 5.56 | 11.78 | 26.08 | 14.07 | 27.05 | 32.38 | 14.21 | 1.50 | 0.52 | 0.41 |
| 1991 | 3.95 | 16.00 | 18.20 | 30.17 | 15.24 | 18.09 | 20.93 | 4.24 | 0.79 | 0.29 |
| 1992 | 0.71 | 16.80 | 33.54 | 18.89 | 16.34 | 6.54 | 5.70 | 5.12 | 1.29 | 0.22 |
| 1993 | 3.57 | 4.75 | 30.81 | 36.46 | 13.22 | 9.90 | 2.13 | 1.75 | 1.17 | 0.36 |
| 1994 | 14.38 | 14.97 | 9.02 | 26.66 | 21.90 | 5.76 | 3.62 | 0.70 | 0.48 | 0.45 |
| 1995 | 1.08 | 29.13 | 24.75 | 8.98 | 23.88 | 17.69 | 3.78 | 1.80 | 0.35 | 0.17 |
| 1996 | 3.72 | 5.42 | 42.58 | 29.44 | 12.89 | 14.62 | 14.02 | 3.80 | 1.04 | 0.18 |
| 1997 | 1.18 | 22.18 | 13.55 | 56.31 | 29.10 | 9.50 | 8.78 | 6.61 | 0.56 | 0.21 |
| 1998 | 8.06 | 5.36 | 29.92 | 16.04 | 61.73 | 28.58 | 6.50 | 5.24 | 3.03 | 0.66 |
| 1999 | 7.39 | 32.98 | 7.01 | 42.25 | 13.00 | 23.66 | 11.12 | 2.35 | 1.32 | 0.70 |
| 2000 | 18.85 | 27.60 | 54.99 | 6.94 | 30.00 | 8.28 | 8.18 | 4.14 | 0.51 | 0.30 |
| 2001 | 12.13 | 21.74 | 36.38 | 38.04 | 4.95 | 15.11 | 3.30 | 1.96 | 0.81 | 0.29 |
| 2002 | 0.91 | 37.85 | 41.22 | 40.13 | 36.25 | 7.09 | 8.32 | 1.49 | 0.72 | 0.30 |
| 2003 | 11.17 | 4.17 | 46.35 | 36.58 | 28.42 | 16.89 | 3.82 | 4.34 | 1.03 | 0.20 |
| 2004 | 6.57 | 24.43 | 7.87 | 61.79 | 35.00 | 24.83 | 14.44 | 2.82 | 2.88 | 0.47 |
| 2005 | 2.56 | 14.54 | 38.70 | 9.68 | 43.57 | 22.97 | 10.84 | 5.77 | 0.93 | 0.92 |
| 2006 | 8.79 | 6.39 | 22.67 | 38.44 | 10.83 | 27.74 | 10.05 | 3.55 | 1.38 | 0.25 |
| 2007 | 5.61 | 18.21 | 8.58 | 21.09 | 27.60 | 9.06 | 9.75 | 5.08 | 2.11 | 0.75 |
| 2008 | 6.40 | 11.77 | 22.08 | 9.31 | 20.43 | 20.40 | 8.10 | 6.63 | 2.47 | 0.60 |
| 2009 | 21.27 | 11.62 | 15.80 | 21.82 | 14.59 | 23.45 | 14.59 | 4.18 | 2.73 | 1.02 |
| 2010 | 18.29 | 20.00 | 18.00 | 17.73 | 23.75 | 13.27 | 16.60 | 8.93 | 2.71 | 1.70 |
| 2011 | 3.57 | 21.49 | 26.63 | 19.90 | 22.48 | 25.32 | 13.51 | 12.31 | 4.55 | 0.91 |
| 2012 | 19.94 | 9.75 | 37.59 | 56.66 | 41.59 | 30.22 | 26.99 | 9.96 | 6.30 | 2.76 |
| 2013 | 10.80 | 31.36 | 17.67 | 43.84 | 46.48 | 25.25 | 16.50 | 13.81 | 6.88 | 3.33 |
| 2014 | 3.31 | 23.97 | 38.00 | 23.48 | 47.17 | 37.60 | 17.31 | 8.18 | 4.26 | 2.22 |
| 2015 | 20.85 | 10.66 | 27.42 | 41.87 | 20.93 | 40.85 | 28.14 | 16.41 | 4.99 | 3.13 |
| 2016 | 31.20 | 29.08 | 14.35 | 36.63 | 53.94 | 27.29 | 36.91 | 18.20 | 6.75 | 2.27 |

Table 7. Icelandic cod in Division 5.a. Survey indices of the fall bottom trawl survey (SMH).

| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 16.54 | 110.43 | 35.40 | 48.20 | 64.15 | 22.57 | 14.85 | 4.85 | 3.21 | 1.76 |
| 1986 | 15.05 | 60.24 | 95.89 | 22.42 | 21.21 | 26.34 | 6.63 | 2.48 | 0.83 | 0.73 |
| 1987 | 3.65 | 28.21 | 103.74 | 81.99 | 21.08 | 12.21 | 12.00 | 2.55 | 0.89 | 0.38 |
| 1988 | 3.44 | 6.96 | 72.09 | 101.49 | 66.56 | 7.81 | 5.90 | 6.29 | 0.58 | 0.24 |
| 1989 | 4.04 | 16.38 | 21.97 | 77.79 | 67.59 | 34.20 | 4.20 | 1.45 | 1.14 | 0.24 |
| 1990 | 5.56 | 11.78 | 26.08 | 14.07 | 27.05 | 32.38 | 14.21 | 1.50 | 0.52 | 0.41 |
| 1991 | 3.95 | 16.00 | 18.20 | 30.17 | 15.24 | 18.09 | 20.93 | 4.24 | 0.79 | 0.29 |
| 1992 | 0.71 | 16.80 | 33.54 | 18.89 | 16.34 | 6.54 | 5.70 | 5.12 | 1.29 | 0.22 |
| 1993 | 3.57 | 4.75 | 30.81 | 36.46 | 13.22 | 9.90 | 2.13 | 1.75 | 1.17 | 0.36 |
| 1994 | 14.38 | 14.97 | 9.02 | 26.66 | 21.90 | 5.76 | 3.62 | 0.70 | 0.48 | 0.45 |
| 1995 | 1.08 | 29.13 | 24.75 | 8.98 | 23.88 | 17.69 | 3.78 | 1.80 | 0.35 | 0.17 |
| 1996 | 3.72 | 5.42 | 42.58 | 29.44 | 12.89 | 14.62 | 14.02 | 3.80 | 1.04 | 0.18 |
| 1997 | 1.18 | 22.18 | 13.55 | 56.31 | 29.10 | 9.50 | 8.78 | 6.61 | 0.56 | 0.21 |
| 1998 | 8.06 | 5.36 | 29.92 | 16.04 | 61.73 | 28.58 | 6.50 | 5.24 | 3.03 | 0.66 |
| 1999 | 7.39 | 32.98 | 7.01 | 42.25 | 13.00 | 23.66 | 11.12 | 2.35 | 1.32 | 0.70 |
| 2000 | 18.85 | 27.60 | 54.99 | 6.94 | 30.00 | 8.28 | 8.18 | 4.14 | 0.51 | 0.30 |
| 2001 | 12.13 | 21.74 | 36.38 | 38.04 | 4.95 | 15.11 | 3.30 | 1.96 | 0.81 | 0.29 |
| 2002 | 0.91 | 37.85 | 41.22 | 40.13 | 36.25 | 7.09 | 8.32 | 1.49 | 0.72 | 0.30 |
| 2003 | 11.17 | 4.17 | 46.35 | 36.58 | 28.42 | 16.89 | 3.82 | 4.34 | 1.03 | 0.20 |
| 2004 | 6.57 | 24.43 | 7.87 | 61.79 | 35.00 | 24.83 | 14.44 | 2.82 | 2.88 | 0.47 |
| 2005 | 2.56 | 14.54 | 38.70 | 9.68 | 43.57 | 22.97 | 10.84 | 5.77 | 0.93 | 0.92 |
| 2006 | 8.79 | 6.39 | 22.67 | 38.44 | 10.83 | 27.74 | 10.05 | 3.55 | 1.38 | 0.25 |
| 2007 | 5.61 | 18.21 | 8.58 | 21.09 | 27.60 | 9.06 | 9.75 | 5.08 | 2.11 | 0.75 |
| 2008 | 6.40 | 11.77 | 22.08 | 9.31 | 20.43 | 20.40 | 8.10 | 6.63 | 2.47 | 0.60 |
| 2009 | 21.27 | 11.62 | 15.80 | 21.82 | 14.59 | 23.45 | 14.59 | 4.18 | 2.73 | 1.02 |
| 2010 | 18.29 | 20.00 | 18.00 | 17.73 | 23.75 | 13.27 | 16.60 | 8.93 | 2.71 | 1.70 |
| 2011 | 3.57 | 21.49 | 26.63 | 19.90 | 22.48 | 25.32 | 13.51 | 12.31 | 4.55 | 0.91 |
| 2012 | 19.94 | 9.75 | 37.59 | 56.66 | 41.59 | 30.22 | 26.99 | 9.96 | 6.30 | 2.76 |
| 2013 | 10.80 | 31.36 | 17.67 | 43.84 | 46.48 | 25.25 | 16.50 | 13.81 | 6.88 | 3.33 |
| 2014 | 3.31 | 23.97 | 38.00 | 23.48 | 47.17 | 37.60 | 17.31 | 8.18 | 4.26 | 2.22 |
| 2015 | 20.85 | 10.66 | 27.42 | 41.87 | 20.93 | 40.85 | 28.14 | 16.41 | 4.99 | 3.13 |
| 2016 | 31.20 | 29.08 | 14.35 | 36.63 | 53.94 | 27.29 | 36.91 | 18.20 | 6.75 | 2.27 |

Table 8. Icelandic cod in Division 5.a. Catch at age residuals from the ADCAM model tuned with the spring (SMB) and the fall (SMH) surveys.

| YEAR | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | -0.12 | -0.21 | 0.08 | 0.12 | 0.21 | -0.12 | -0.16 | 0.13 | -0.10 | -0.45 | -0.21 | 0.00 |
| 1956 | -0.03 | -0.05 | 0.03 | 0.00 | -0.13 | -0.20 | -0.01 | 0.01 | 0.17 | 0.09 | 0.23 | 0.22 |
| 1957 | 0.09 | 0.02 | -0.01 | 0.17 | -0.13 | 0.09 | 0.06 | -0.15 | -0.10 | -0.12 | -0.38 | 0.52 |
| 1958 | 0.15 | 0.18 | -0.26 | -0.07 | 0.06 | 0.08 | 0.13 | -0.23 | 0.23 | 0.00 | -0.23 | 0.39 |
| 1959 | -0.22 | 0.21 | 0.26 | -0.24 | -0.22 | -0.06 | -0.06 | 0.28 | -0.27 | 0.38 | -0.23 | -0.42 |
| 1960 | 0.10 | -0.36 | 0.14 | 0.19 | 0.06 | 0.07 | -0.03 | -0.11 | -0.04 | 0.03 | -0.64 | 0.89 |
| 1961 | 0.05 | 0.04 | -0.40 | 0.12 | -0.02 | 0.27 | 0.20 | -0.14 | 0.09 | -0.19 | -0.97 | 0.82 |
| 1962 | 0.09 | -0.01 | 0.13 | -0.24 | 0.12 | -0.29 | 0.09 | 0.26 | -0.06 | 0.04 | -0.40 | 0.69 |
| 1963 | -0.06 | 0.30 | -0.18 | 0.01 | -0.03 | -0.07 | -0.38 | 0.21 | 0.35 | 0.06 | 0.08 | -0.62 |
| 1964 | -0.13 | -0.02 | 0.13 | -0.25 | -0.12 | 0.38 | -0.10 | -0.45 | -0.01 | 0.27 | -0.15 | 0.00 |
| 1965 | -0.03 | -0.11 | 0.09 | 0.16 | -0.13 | 0.05 | 0.47 | -0.48 | -0.06 | -0.51 | -0.36 | 0.63 |
| 1966 | -0.04 | -0.04 | -0.18 | 0.10 | -0.07 | 0.12 | -0.35 | 0.59 | -0.83 | 0.28 | 0.01 | 1.05 |
| 1967 | 0.19 | -0.13 | 0.02 | -0.20 | 0.02 | -0.37 | 0.49 | 0.05 | 0.67 | -0.72 | -0.83 | -0.19 |
| 1968 | 0.04 | -0.02 | -0.27 | -0.12 | 0.23 | 0.16 | -0.42 | 0.37 | -0.12 | 0.60 | -0.66 | 0.64 |
| 1969 | -0.09 | -0.03 | 0.15 | -0.01 | 0.05 | -0.15 | -0.33 | -0.25 | -0.04 | -0.26 | -0.81 | -0.16 |
| 1970 | -0.10 | 0.14 | -0.05 | -0.14 | 0.05 | -0.16 | 0.48 | -0.58 | -0.12 | 0.25 | 0.30 | 0.44 |
| 1971 | -0.10 | 0.07 | 0.09 | 0.18 | -0.18 | 0.28 | -0.17 | 0.06 | -0.45 | -0.02 | 0.13 | 0.35 |
| 1972 | -0.17 | -0.1 | 0.07 | -0.03 | 0.12 | -0.0 | -0.10 | 0.29 | -0.07 | 0.17 | 0.53 | -2.78 |
| 1973 | 0.28 | -0.02 | -0.10 | 0.03 | 0.00 | -0.2 | 0.09 | 0.17 | 0.16 | -0.19 | -1.25 | -2.12 |
| 1974 | -0.16 | 0.21 | -0.02 | -0.18 | -0.01 | 0.00 | -0.22 | 0.29 | 0.01 | 0.19 | -0.43 | 0.78 |
| 1975 | 0.19 | -0.07 | 0.04 | -0.05 | 0.03 | -0.15 | -0.21 | 0.00 | 0.41 | -0.01 | -0.12 | 0.06 |
| 1976 | 0.10 | 0.00 | -0.17 | 0.08 | -0.09 | 0.25 | -0.16 | -0.15 | 0.06 | 0.27 | -0.23 | 0.20 |
| 1977 | -0.40 | -0.06 | 0.05 | -0.09 | 0.13 | 0.05 | 0.31 | 0.03 | -0.70 | -0.48 | -1.22 | -2.53 |
| 1978 | 0.08 | -0.01 | 0.04 | -0.10 | 0.04 | -0.21 | 0.12 | -0.19 | 0.01 | -0.05 | 0.54 | 1.17 |
| 1979 | 0.16 | 0.09 | -0.22 | 0.10 | -0.05 | 0.03 | -0.31 | -0.08 | 0.04 | -0.14 | 0.42 | -0.22 |
| 1980 | 0.21 | 0.01 | 0.08 | 0.06 | -0.01 | -0.09 | 0.12 | -0.49 | 0.29 | 0.10 | 0.17 | -1.09 |
| 1981 | -0.30 | -0.21 | 0.08 | -0.14 | 0.07 | 0.09 | 0.02 | 0.33 | -0.08 | 0.60 | 0.00 | 1.16 |
| 1982 | 0.01 | 0.15 | 0.07 | -0.05 | -0.22 | 0.19 | 0.18 | 0.14 | -0.23 | -0.87 | 0.06 | -0.88 |
| 1983 | -0.32 | -0.36 | 0.11 | 0.14 | 0.04 | 0.01 | -0.04 | -0.03 | 0.00 | 0.37 | -0.19 | 0.55 |
| 1984 | 0.35 | 0.03 | -0.06 | -0.05 | -0.10 | 0.00 | 0.05 | -0.14 | -0.36 | 0.17 | 0.72 | 0.07 |
| 1985 | 0.04 | 0.18 | -0.10 | 0.12 | -0.10 | -0.02 | -0.14 | 0.13 | 0.03 | -0.34 | 0.49 | 0.44 |
| 1986 | 0.14 | -0.12 | 0.01 | -0.01 | 0.18 | -0.05 | 0.12 | -0.21 | 0.08 | 0.06 | -0.57 | 0.15 |
| 1987 | -0.15 | 0.12 | 0.01 | -0.16 | 0.07 | 0.04 | -0.03 | 0.11 | -0.38 | -0.11 | 0.14 | -0.34 |
| 1988 | -0.09 | -0.06 | -0.06 | 0.13 | -0.08 | 0.07 | 0.16 | 0.03 | 0.47 | 0.02 | 0.54 | 0.04 |
| 1989 | -0.21 | 0.04 | 0.14 | -0.08 | 0.00 | -0.15 | -0.32 | -0.09 | -0.03 | 0.52 | -0.02 | -1.51 |
| 1990 | 0.00 | -0.14 | -0.11 | 0.00 | 0.04 | 0.09 | -0.08 | -0.23 | 0.29 | 0.13 | -0.20 | 0.00 |
| 1991 | 0.07 | 0.04 | -0.13 | -0.07 | 0.10 | -0.08 | 0.12 | -0.07 | -0.31 | 0.42 | -0.55 | 0.04 |
| 1992 | -0.23 | 0.08 | 0.04 | 0.03 | 0.10 | 0.00 | -0.05 | -0.07 | -0.75 | -0.76 | -0.57 | -0.25 |
| 1993 | 0.25 | 0.04 | -0.20 | -0.06 | -0.07 | -0.12 | 0.07 | 0.48 | 0.49 | -0.21 | -1.00 | 0.29 |
| 1994 | 0.03 | 0.25 | -0.14 | -0.19 | -0.04 | 0.07 | -0.19 | -0.14 | 0.42 | 0.52 | 0.50 | -0.56 |
| 1995 | 0.27 | -0.04 | 0.08 | -0.04 | -0.04 | -0.12 | -0.13 | -0.29 | -0.21 | 0.75 | 1.12 | 0.48 |
| 1996 | 0.00 | -0.05 | -0.18 | 0.08 | 0.04 | 0.02 | 0.13 | 0.18 | -0.38 | -0.37 | 0.64 | -0.16 |


| YEAR | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | -0.16 | 0.03 | -0.03 | -0.13 | -0.09 | 0.21 | 0.18 | 0.26 | 0.42 | -0.70 | -0.19 | 0.08 |
| 1998 | -0.19 | -0.17 | 0.07 | 0.07 | 0.02 | -0.16 | 0.25 | 0.06 | 0.10 | 0.31 | 0.17 | -0.85 |
| 1999 | -0.11 | 0.04 | 0.04 | 0.03 | 0.09 | -0.04 | -0.23 | -0.18 | -0.25 | -0.39 | -0.50 | -1.09 |
| 2000 | 0.14 | -0.23 | 0.12 | -0.04 | 0.01 | 0.11 | 0.04 | -0.12 | -0.01 | 0.13 | -0.21 | -0.35 |
| 2001 | 0.24 | 0.22 | -0.19 | 0.02 | 0.04 | -0.25 | 0.07 | 0.33 | -0.14 | 0.19 | -0.71 | 0.80 |
| 2002 | -0.10 | 0.05 | 0.03 | -0.04 | 0.01 | 0.03 | -0.08 | 0.25 | 0.31 | -0.13 | 0.09 | -0.98 |
| 2003 | -0.07 | 0.03 | 0.00 | -0.11 | 0.12 | 0.07 | 0.21 | -0.18 | 0.10 | 0.26 | 0.08 | 0.68 |
| 2004 | -0.26 | 0.06 | 0.08 | -0.05 | -0.05 | 0.25 | -0.04 | 0.23 | -0.34 | 0.23 | 0.27 | -0.40 |
| 2005 | 0.17 | -0.26 | 0.13 | -0.05 | -0.11 | -0.08 | 0.31 | -0.01 | 0.41 | 0.21 | 0.32 | -0.30 |
| 2006 | -0.06 | 0.04 | -0.10 | 0.06 | 0.06 | -0.07 | -0.08 | 0.21 | -0.22 | -0.22 | -0.64 | -0.23 |
| 2007 | -0.15 | 0.14 | -0.05 | 0.02 | -0.13 | 0.08 | 0.02 | 0.18 | 0.78 | -0.22 | 1.15 | -0.95 |
| 2008 | 0.12 | -0.18 | 0.07 | -0.15 | 0.05 | -0.19 | 0.04 | 0.14 | -0.08 | 0.46 | 0.25 | -0.68 |
| 2009 | 0.08 | -0.11 | 0.06 | 0.16 | -0.02 | 0.23 | -0.20 | -0.24 | -0.02 | -0.32 | 0.23 | -0.43 |
| 2010 | 0.08 | 0.03 | -0.15 | 0.04 | 0.05 | -0.07 | 0.20 | -0.20 | -0.17 | 0.19 | 0.34 | 0.67 |
| 2011 | -0.02 | -0.03 | 0.08 | 0.01 | -0.02 | -0.01 | -0.11 | 0.10 | -0.13 | -0.12 | -0.20 | -0.42 |
| 2012 | -0.15 | 0.03 | 0.00 | -0.03 | 0.00 | 0.17 | 0.01 | -0.25 | -0.02 | -0.09 | 0.09 | -0.23 |
| 2013 | 0.26 | -0.03 | 0.02 | 0.03 | -0.07 | -0.05 | 0.16 | -0.01 | -0.19 | 0.22 | 0.11 | -0.44 |
| 2014 | -0.11 | 0.03 | 0.01 | 0.00 | 0.02 | -0.06 | -0.06 | 0.12 | -0.02 | -0.11 | 0.40 | 0.09 |
| 2015 | 0.02 | 0.06 | 0.03 | 0.04 | -0.10 | 0.03 | -0.06 | -0.07 | 0.41 | -0.11 | -0.37 | -0.35 |

Table 9. Icelandic cod in Division 5.a. Spring survey (SMB) at age residuals from the ADCAM model, assessment tuned with both the spring and the fall survey.

| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | -0.48 | 0.07 | 0.24 | 0.46 | 0.14 | 0.26 | 0.40 | 0.18 | 0.31 | 0.68 |
| 1986 | 0.44 | -0.03 | -0.38 | -0.22 | -0.07 | 0.00 | -0.17 | -0.28 | -0.25 | -0.03 |
| 1987 | 0.65 | 0.03 | 0.14 | -0.44 | -0.02 | -0.07 | 0.04 | -0.09 | -0.10 | 0.00 |
| 1988 | -0.19 | 0.04 | 0.52 | 0.17 | -0.12 | -0.34 | 0.08 | 0.48 | -0.12 | -0.09 |
| 1989 | 0.39 | 0.08 | 0.55 | 0.57 | 0.24 | 0.18 | -0.13 | -0.11 | 0.21 | 0.11 |
| 1990 | -0.47 | 0.13 | 0.09 | 0.07 | -0.15 | -0.15 | 0.07 | -0.16 | -0.05 | 0.16 |
| 1991 | -0.16 | -0.43 | 0.12 | 0.17 | 0.27 | 0.04 | 0.12 | -0.16 | 0.22 | 0.20 |
| 1992 | -0.24 | 0.04 | -0.18 | 0.13 | -0.08 | -0.13 | -0.15 | -0.15 | -0.12 | 0.00 |
| 1993 | -0.51 | -0.03 | 0.20 | -0.04 | 0.06 | -0.04 | -0.22 | -0.16 | -0.23 | -0.22 |
| 1994 | 0.54 | -0.24 | 0.04 | 0.13 | -0.19 | -0.32 | -0.17 | -0.23 | -0.20 | -0.06 |
| 1995 | -0.23 | 0.16 | -0.21 | -0.04 | 0.18 | -0.02 | -0.23 | -0.10 | -0.07 | -0.21 |
| 1996 | -0.63 | -0.11 | 0.11 | -0.11 | 0.21 | -0.05 | 0.24 | 0.39 | 0.19 | 0.05 |
| 1997 | 0.19 | -0.03 | 0.14 | 0.29 | -0.03 | -0.05 | -0.05 | 0.24 | -0.36 | -0.30 |
| 1998 | -0.10 | 0.13 | -0.17 | 0.13 | 0.51 | 0.29 | 0.08 | 0.19 | 0.42 | 0.49 |
| 1999 | -0.04 | 0.20 | -0.02 | 0.06 | -0.05 | 0.07 | 0.01 | -0.04 | -0.04 | 0.11 |
| 2000 | 0.90 | 0.14 | 0.30 | -0.16 | -0.09 | -0.23 | -0.23 | -0.04 | -0.28 | -0.28 |
| 2001 | 0.20 | 0.04 | 0.02 | -0.08 | -0.45 | -0.24 | -0.41 | -0.60 | -0.39 | 0.14 |
| 2002 | -0.15 | 0.27 | 0.16 | 0.08 | 0.06 | -0.15 | -0.20 | -0.32 | -0.46 | -0.22 |
| 2003 | 0.03 | -0.12 | 0.06 | -0.02 | -0.11 | -0.21 | -0.21 | -0.08 | 0.14 | -0.56 |
| 2004 | -0.07 | 0.19 | -0.10 | 0.29 | 0.12 | 0.22 | 0.18 | 0.14 | 0.44 | 0.30 |
| 2005 | -0.14 | 0.09 | 0.22 | -0.11 | 0.09 | 0.10 | -0.02 | 0.04 | 0.05 | 0.31 |
| 2006 | 0.21 | -0.05 | -0.01 | 0.08 | -0.08 | 0.15 | -0.12 | -0.33 | -0.32 | -0.17 |
| 2007 | 0.03 | 0.17 | -0.31 | -0.22 | -0.16 | -0.18 | -0.32 | -0.07 | 0.04 | -0.05 |
| 2008 | -0.02 | 0.00 | -0.06 | -0.41 | -0.27 | -0.11 | 0.11 | -0.05 | 0.10 | -0.17 |
| 2009 | 0.41 | -0.11 | -0.16 | -0.21 | -0.17 | -0.09 | -0.08 | 0.03 | -0.19 | -0.09 |
| 2010 | 0.12 | -0.16 | -0.19 | -0.24 | -0.19 | -0.20 | -0.08 | -0.04 | 0.36 | 0.05 |
| 2011 | -0.45 | -0.20 | -0.35 | -0.28 | -0.11 | 0.06 | 0.10 | 0.09 | -0.02 | -0.08 |
| 2012 | 0.17 | -0.14 | -0.12 | 0.20 | 0.33 | 0.30 | 0.41 | 0.27 | 0.13 | 0.13 |
| 2013 | -0.14 | 0.10 | -0.14 | -0.11 | 0.05 | 0.06 | 0.03 | 0.24 | 0.54 | 0.03 |
| 2014 | -0.38 | 0.01 | -0.16 | -0.05 | -0.01 | 0.17 | -0.01 | -0.17 | -0.25 | 0.01 |
| 2015 | -0.12 | -0.05 | -0.25 | -0.18 | -0.21 | 0.20 | 0.19 | 0.39 | -0.02 | 0.02 |
| 2016 | 0.28 | -0.19 | -0.19 | -0.10 | 0.11 | 0.18 | 0.39 | 0.22 | 0.17 | -0.21 |

Table 10. Icelandic cod in Division 5.a. fall survey (SMH) at age residuals from the ADCAM model, assessment tuned with both the spring and the fall survey.

| YEAR | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 0.01 | -0.09 | -0.02 | -0.20 | -0.02 | -0.08 | 0.16 | 0.19 | -0.18 | -0.04 |
| 1997 | -0.12 | 0.11 | -0.03 | 0.24 | 0.04 | -0.17 | -0.14 | -0.04 | -0.34 | -0.05 |
| 1998 | -0.24 | -0.02 | -0.20 | 0.02 | -0.05 | 0.34 | 0.50 | 0.11 | 0.25 | 0.05 |
| 1999 | 0.23 | -0.09 | 0.12 | 0.10 | 0.06 | -0.02 | -0.12 | -0.31 | -0.37 | 0.09 |
| 2000 | -0.28 | -0.08 | -0.27 | -0.08 | -0.24 | -0.24 | -0.41 | -0.35 | -0.02 | 0.16 |
| 2001 | -0.16 | -0.15 | 0.03 | -0.02 | -0.23 | -0.27 | -0.26 | -0.53 | -0.61 | -0.38 |
| 2002 | -0.13 | -0.20 | -0.13 | 0.15 | -0.01 | 0.11 | -0.02 | 0.00 | -0.04 | -0.43 |
| 2003 | -0.12 | -0.11 | 0.08 | -0.15 | -0.12 | -0.16 | -0.13 | 0.07 | -0.06 | -0.44 |
| 2004 | -0.12 | 0.15 | 0.10 | 0.12 | 0.16 | 0.09 | 0.22 | 0.33 | 0.48 | 0.21 |
| 2005 | 0.10 | -0.08 | 0.09 | 0.08 | 0.24 | -0.02 | -0.28 | -0.28 | -0.21 | -0.06 |
| 2006 | 0.09 | -0.08 | 0.09 | 0.09 | 0.07 | 0.04 | 0.02 | -0.21 | -0.06 | -0.04 |
| 2007 | 0.14 | 0.00 | -0.34 | -0.27 | -0.11 | -0.03 | -0.20 | 0.02 | -0.27 | 0.08 |
| 2008 | 0.28 | 0.27 | 0.05 | -0.14 | 0.09 | 0.22 | 0.27 | 0.25 | 0.05 | 0.35 |
| 2009 | -0.08 | -0.08 | 0.08 | 0.08 | 0.13 | 0.04 | 0.13 | 0.23 | 0.26 | 0.16 |
| 2010 | 0.19 | 0.11 | 0.14 | 0.11 | 0.09 | -0.02 | 0.10 | 0.21 | 0.52 | 0.09 |
| 2011 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2012 | -0.20 | 0.13 | -0.03 | -0.22 | -0.21 | -0.16 | 0.05 | 0.21 | -0.09 | -0.02 |
| 2013 | -0.09 | 0.02 | 0.06 | -0.05 | -0.03 | -0.07 | -0.07 | 0.00 | 0.19 | -0.06 |
| 2014 | 0.16 | 0.05 | -0.04 | -0.01 | -0.02 | 0.08 | -0.06 | 0.08 | 0.18 | 0.59 |
| 2015 | 0.32 | 0.10 | 0.22 | 0.18 | 0.15 | 0.30 | 0.24 | -0.08 | 0.06 | -0.43 |
|  |  |  |  |  |  |  |  |  |  |  |

Table 11. Icelandic cod in Division 5.a. Estimates of fishing mortality 1955-2015 based on ACAM using catch at age and spring and fall bottom survey indices.

| YEAR | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 0.040 | 0.170 | 0.252 | 0.273 | 0.301 | 0.302 | 0.280 | 0.319 | 0.319 | 0.304 | 0.319 | 0.319 |
| 1956 | 0.051 | 0.182 | 0.249 | 0.258 | 0.289 | 0.302 | 0.292 | 0.338 | 0.350 | 0.329 | 0.326 | 0.326 |
| 1957 | 0.081 | 0.215 | 0.273 | 0.271 | 0.300 | 0.326 | 0.325 | 0.359 | 0.359 | 0.325 | 0.291 | 0.291 |
| 1958 | 0.114 | 0.248 | 0.302 | 0.290 | 0.323 | 0.370 | 0.395 | 0.434 | 0.437 | 0.377 | 0.315 | 0.315 |
| 1959 | 0.091 | 0.233 | 0.282 | 0.256 | 0.298 | 0.341 | 0.350 | 0.396 | 0.377 | 0.314 | 0.223 | 0.223 |
| 1960 | 0.101 | 0.233 | 0.295 | 0.292 | 0.338 | 0.397 | 0.427 | 0.473 | 0.470 | 0.378 | 0.263 | 0.263 |
| 1961 | 0.094 | 0.225 | 0.259 | 0.262 | 0.333 | 0.398 | 0.418 | 0.457 | 0.436 | 0.343 | 0.221 | 0.221 |
| 1962 | 0.112 | 0.248 | 0.282 | 0.264 | 0.346 | 0.424 | 0.467 | 0.511 | 0.484 | 0.372 | 0.233 | 0.233 |
| 1963 | 0.130 | 0.283 | 0.328 | 0.309 | 0.383 | 0.492 | 0.587 | 0.644 | 0.622 | 0.456 | 0.279 | 0.279 |
| 1964 | 0.126 | 0.290 | 0.372 | 0.360 | 0.434 | 0.570 | 0.740 | 0.808 | 0.830 | 0.601 | 0.379 | 0.379 |
| 1965 | 0.121 | 0.284 | 0.384 | 0.403 | 0.471 | 0.601 | 0.743 | 0.845 | 0.874 | 0.645 | 0.415 | 0.415 |
| 1966 | 0.094 | 0.253 | 0.341 | 0.381 | 0.49 | 0.621 | 0.780 | 0.913 | 1.001 | 0.773 | 0.518 | 0.518 |
| 1967 | 0.07 | 0.22 | 0.303 | 0.338 | 0.48 | 0.60 | 0.7 | 0.875 | 0.923 | 0.712 | 0.446 | 0.446 |
| 1968 | 0.077 | 0.247 | 0.342 | 0.405 | 0.575 | 0.765 | 1.035 | 1.196 | 1.351 | 1.063 | 0.714 | 0.714 |
| 1969 | 0.056 | 0.232 | 0.322 | 0.354 | 0.504 | 0.608 | 0.718 | 0.834 | 0.864 | 0.701 | 0.428 | 0.428 |
| 1970 | 0.069 | 0.269 | 0.389 | 0.426 | 0.551 | 0.649 | 0.759 | 0.888 | 0.942 | 0.784 | 0.494 | 0.494 |
| 1971 | 0.088 | 0.309 | 0.478 | 0.532 | 0.620 | 0.716 | 0.799 | 0.953 | 1.025 | 0.861 | 0.553 | 0.553 |
| 1972 | 0.088 | 0.302 | 0.479 | 0.553 | 0.649 | 0.729 | 0.790 | 0.955 | 1.050 | 0.889 | 0.572 | 0.572 |
| 1973 | 0.119 | 0.320 | 0.488 | 0.564 | 0.66 | 0.753 | 0.798 | 0.949 | 1.032 | 0.877 | 0.560 | 0.560 |
| 1974 | 0.113 | 0.324 | 0.498 | 0.574 | 0.69 | 0.830 | 0.918 | 1.050 | 1.166 | 0.996 | 0.656 | 0.656 |
| 1975 | 0.108 | 0.309 | 0.501 | 0.600 | 0.72 | 0.881 | 1.016 | 1.116 | 1.231 | 1.060 | 0.717 | 0.717 |
| 1976 | 0.066 | 0.258 | 0.427 | 0.550 | 0.69 | 0.848 | 0.939 | 0.993 | 1.036 | 0.899 | 0.600 | 0.600 |
| 1977 | 0.030 | 0.195 | 0.329 | 0.427 | 0.60 | 0.717 | 0.719 | 0.725 | 0.676 | 0.593 | 0.370 | 0.370 |
| 1978 | 0.027 | 0.174 | 0.281 | 0.354 | 0.524 | 0.599 | 0.540 | 0.538 | 0.470 | 0.423 | 0.255 | 0.255 |
| 1979 | 0.028 | 0.171 | 0.274 | 0.344 | 0.501 | 0.564 | 0.491 | 0.482 | 0.407 | 0.370 | 0.223 | 0.223 |
| 1980 | 0.028 | 0.175 | 0.306 | 0.386 | 0.537 | 0.618 | 0.553 | 0.537 | 0.457 | 0.416 | 0.264 | 0.264 |
| 1981 | 0.023 | 0.176 | 0.353 | 0.488 | 0.647 | 0.816 | 0.843 | 0.804 | 0.730 | 0.652 | 0.468 | 0.468 |
| 1982 | 0.028 | 0.192 | 0.395 | 0.557 | 0.697 | 0.894 | 0.949 | 0.852 | 0.723 | 0.629 | 0.456 | 0.456 |
| 1983 | 0.023 | 0.179 | 0.377 | 0.554 | 0.704 | 0.877 | 0.904 | 0.834 | 0.708 | 0.626 | 0.463 | 0.463 |
| 1984 | 0.039 | 0.200 | 0.378 | 0.530 | 0.673 | 0.801 | 0.743 | 0.686 | 0.573 | 0.518 | 0.378 | 0.378 |
| 1985 | 0.050 | 0.230 | 0.422 | 0.577 | 0.712 | 0.827 | 0.754 | 0.681 | 0.568 | 0.514 | 0.378 | 0.378 |
| 1986 | 0.061 | 0.261 | 0.517 | 0.713 | 0.822 | 0.948 | 0.861 | 0.747 | 0.629 | 0.559 | 0.415 | 0.415 |
| 1987 | 0.055 | 0.272 | 0.555 | 0.816 | 0.903 | 1.052 | 0.976 | 0.823 | 0.705 | 0.624 | 0.477 | 0.477 |
| 1988 | 0.047 | 0.258 | 0.523 | 0.794 | 0.918 | 1.095 | 1.059 | 0.909 | 0.822 | 0.734 | 0.588 | 0.588 |
| 1989 | 0.041 | 0.241 | 0.463 | 0.653 | 0.790 | 0.885 | 0.779 | 0.688 | 0.599 | 0.543 | 0.407 | 0.407 |
| 1990 | 0.050 | 0.250 | 0.471 | 0.662 | 0.785 | 0.849 | 0.730 | 0.656 | 0.570 | 0.515 | 0.383 | 0.383 |
| 1991 | 0.086 | 0.301 | 0.565 | 0.811 | 0.879 | 0.936 | 0.820 | 0.732 | 0.650 | 0.580 | 0.443 | 0.443 |
| 1992 | 0.102 | 0.320 | 0.598 | 0.868 | 0.918 | 0.990 | 0.865 | 0.759 | 0.673 | 0.593 | 0.457 | 0.457 |
| 1993 | 0.138 | 0.313 | 0.554 | 0.801 | 0.882 | 1.015 | 0.988 | 0.878 | 0.810 | 0.706 | 0.568 | 0.568 |
| 1994 | 0.088 | 0.241 | 0.383 | 0.530 | 0.671 | 0.752 | 0.688 | 0.651 | 0.579 | 0.518 | 0.397 | 0.397 |
| 1995 | 0.061 | 0.195 | 0.318 | 0.421 | 0.564 | 0.614 | 0.536 | 0.531 | 0.463 | 0.421 | 0.315 | 0.315 |
| 1996 | 0.036 | 0.160 | 0.282 | 0.410 | 0.553 | 0.614 | 0.556 | 0.554 | 0.484 | 0.433 | 0.328 | 0.328 |


| YEAR | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 0.025 | 0.144 | 0.274 | 0.419 | 0.577 | 0.657 | 0.632 | 0.628 | 0.559 | 0.488 | 0.380 | 0.380 |
| 1998 | 0.029 | 0.152 | 0.328 | 0.516 | 0.655 | 0.762 | 0.774 | 0.752 | 0.689 | 0.590 | 0.478 | 0.478 |
| 1999 | 0.045 | 0.175 | 0.387 | 0.640 | 0.732 | 0.840 | 0.862 | 0.808 | 0.740 | 0.625 | 0.514 | 0.514 |
| 2000 | 0.060 | 0.179 | 0.384 | 0.609 | 0.727 | 0.846 | 0.879 | 0.839 | 0.777 | 0.656 | 0.551 | 0.551 |
| 2001 | 0.075 | 0.190 | 0.380 | 0.573 | 0.682 | 0.823 | 0.910 | 0.894 | 0.849 | 0.714 | 0.618 | 0.618 |
| 2002 | 0.047 | 0.163 | 0.335 | 0.483 | 0.593 | 0.707 | 0.789 | 0.805 | 0.759 | 0.641 | 0.550 | 0.550 |
| 2003 | 0.037 | 0.150 | 0.328 | 0.489 | 0.571 | 0.668 | 0.719 | 0.745 | 0.702 | 0.596 | 0.513 | 0.513 |
| 2004 | 0.032 | 0.141 | 0.324 | 0.515 | 0.578 | 0.676 | 0.726 | 0.745 | 0.708 | 0.598 | 0.522 | 0.522 |
| 2005 | 0.032 | 0.126 | 0.289 | 0.472 | 0.546 | 0.643 | 0.686 | 0.704 | 0.675 | 0.567 | 0.494 | 0.494 |
| 2006 | 0.030 | 0.120 | 0.264 | 0.454 | 0.530 | 0.632 | 0.671 | 0.671 | 0.634 | 0.523 | 0.452 | 0.452 |
| 2007 | 0.027 | 0.108 | 0.229 | 0.380 | 0.484 | 0.597 | 0.652 | 0.651 | 0.620 | 0.508 | 0.442 | 0.442 |
| 2008 | 0.023 | 0.090 | 0.181 | 0.294 | 0.401 | 0.484 | 0.491 | 0.485 | 0.434 | 0.359 | 0.292 | 0.292 |
| 2009 | 0.030 | 0.095 | 0.186 | 0.303 | 0.400 | 0.480 | 0.472 | 0.446 | 0.387 | 0.316 | 0.250 | 0.250 |
| 2010 | 0.028 | 0.090 | 0.165 | 0.253 | 0.350 | 0.411 | 0.376 | 0.354 | 0.294 | 0.243 | 0.182 | 0.182 |
| 2011 | 0.027 | 0.089 | 0.162 | 0.240 | 0.324 | 0.375 | 0.328 | 0.307 | 0.246 | 0.200 | 0.143 | 0.143 |
| 2012 | 0.029 | 0.092 | 0.166 | 0.244 | 0.322 | 0.370 | 0.322 | 0.299 | 0.240 | 0.194 | 0.137 | 0.137 |
| 2013 | 0.040 | 0.101 | 0.177 | 0.254 | 0.327 | 0.381 | 0.335 | 0.316 | 0.262 | 0.209 | 0.149 | 0.149 |
| 2014 | 0.031 | 0.096 | 0.164 | 0.232 | 0.307 | 0.363 | 0.327 | 0.318 | 0.276 | 0.214 | 0.155 | 0.155 |
| 2015 | 0.034 | 0.099 | 0.164 | 0.221 | 0.291 | 0.346 | 0.308 | 0.306 | 0.268 | 0.199 | 0.135 | 0.135 |

Table 12. Icelandic cod in Division 5.a. Estimates of numbers at age in the stock 1955-2016 based on ACAM using catch at age and spring and fall bottom survey indices.

| YeAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 254.563 | 186.621 | 152.016 | 217.665 | 212.115 | 115.674 | 36.147 | 24.683 | 13.037 | 88.714 | 9.306 | 7.913 | 8.261 | 2.680 |
| 1956 | 329.670 | 208.419 | 152.792 | 119.578 | 150.337 | 134.944 | 72.052 | 21.901 | 14.940 | 8.063 | 52.776 | 5.536 | 4.781 | 4.914 |
| 1957 | 431.693 | 269.911 | 170.639 | 118.882 | 81.607 | 95.953 | 85.376 | 44.193 | 13.258 | 9.131 | 4.709 | 30.454 | 3.262 | 2.825 |
| 1958 | 230.244 | 353.441 | 220.984 | 128.816 | 78.524 | 50.834 | 59.908 | 51.798 | 35.301 | 7.843 | 5.221 | 2.693 | 18.015 | 1.997 |
| 1959 | 288.032 | 188.508 | 289.373 | 161.372 | 82.270 | 47.548 | 31.134 | 35.512 | 51.461 | 19.463 | 4.159 | 2.761 | 1.512 | 10.766 |
| 1960 | 192.352 | 235.821 | 154.337 | 216.341 | 104.638 | 50.789 | 30.125 | 18.915 | 20.681 | 37.649 | 10.729 | 2.335 | 1.651 | 0.991 |
| 1961 | 264.814 | 157.485 | 193.074 | 114.255 | 140.289 | 63.794 | 31.062 | 17.596 | 10.412 | 11.044 | 19.204 | 5.492 | 1.309 | 1.040 |
| 1962 | 304.180 | 216.812 | 128.938 | 143.909 | 74.689 | 88.615 | 40.205 | 18.222 | 23.625 | 5.614 | 5.728 | 10.169 | 3.190 | 0.860 |
| 1963 | 322.813 | 249.041 | 177.510 | 94.417 | 91.934 | 46.136 | 55.737 | 23.281 | 9.767 | 12.131 | 2.758 | 2.890 | 5.737 | 2.069 |
| 1964 | 341.867 | 264.297 | 203.898 | 127.656 | 58.249 | 54.216 | 27.732 | 31.116 | 11.656 | 4.449 | 5.217 | 1.213 | 1.500 | 3.554 |
| 1965 | 477.837 | 279.897 | 216.388 | 147.222 | 78.184 | 32.876 | 30.956 | 14.704 | 14.408 | 4.554 | 1.624 | 1.862 | 0.544 | 0.840 |
| 1966 | 256.297 | 391.219 | 229.160 | 157.034 | 90.750 | 43.587 | 17.993 | 15.828 | 6.599 | 5.609 | 1.601 | 0.555 | 0.800 | 0.294 |
| 1967 | 369.071 | 209.839 | 320.303 | 170.783 | 99.797 | 52.835 | 24.373 | 9.020 | 6.961 | 2.477 | 1.843 | 0.482 | 0.210 | 0.390 |
| 1968 | 269.152 | 302.170 | 171.801 | 242.900 | 111.218 | 60.347 | 30.858 | 12.305 | 4.016 | 2.696 | 0.845 | 0.600 | 0.194 | 0.110 |
| 1969 | 281.354 | 220.363 | 247.395 | 130.267 | 155.406 | 64.703 | 32.947 | 41.239 | 4.690 | 1.168 | 0.667 | 0.179 | 0.170 | 0.078 |
| 1970 | 207.686 | 230.354 | 180.418 | 191.546 | 84.546 | 92.175 | 37.184 | 32.930 | 18.385 | 1.872 | 0.415 | 0.230 | 0.073 | 0.091 |
| 1971 | 407.528 | 170.039 | 188.597 | 137.914 | 119.777 | 46.911 | 49.309 | 17.552 | 14.083 | 7.045 | 0.631 | 0.132 | 0.086 | 0.036 |
| 1972 | 267.009 | 333.655 | 139.216 | 141.360 | 82.919 | 60.800 | 22.564 | 21.724 | 23.321 | 5.188 | 2.225 | 0.185 | 0.046 | 0.041 |
| 1973 | 389.158 | 218.609 | 273.174 | 104.426 | 85.604 | 42.046 | 28.639 | 9.655 | 8.583 | 8.663 | 1.634 | 0.637 | 0.062 | 0.021 |
| 1974 | 548.519 | 318.616 | 178.982 | 198.643 | 62.064 | 43.028 | 19.591 | 12.035 | 3.722 | 3.165 | 2.745 | 0.477 | 0.217 | 0.029 |
| 1975 | 213.808 | 449.089 | 260.861 | 130.851 | 117.590 | 30.871 | 19.836 | 7.978 | 4.296 | 1.216 | 0.907 | 0.700 | 0.144 | 0.092 |
| 1976 | 339.523 | 175.051 | 367.683 | 191.686 | 78.627 | 58.308 | 13.877 | 7.900 | 2.705 | 1.273 | 0.326 | 0.217 | 0.199 | 0.058 |
| 1977 | 363.166 | 277.978 | 143.320 | 281.726 | 121.245 | 41.995 | 27.530 | 5.681 | 2.770 | 0.866 | 0.386 | 0.095 | 0.072 | 0.089 |
| 1978 | 208.898 | 297.335 | 227.589 | 113.822 | 189.765 | 71.428 | 22.428 | 12.279 | 2.272 | 1.105 | 0.344 | 0.161 | 0.043 | 0.041 |


| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 209.310 | 171.031 | 243.438 | 181.328 | 78.320 | 117.303 | 41.057 | 10.877 | 5.522 | 1.083 | 0.528 | 0.176 | 0.086 | 0.027 |
| 1980 | 196.438 | 171.368 | 140.028 | 193.763 | 125.168 | 48.747 | 71.952 | 20.359 | 5.065 | 2.766 | 0.548 | 0.288 | 0.099 | 0.056 |
| 1981 | 347.037 | 160.830 | 140.305 | 111.465 | 133.194 | 75.479 | 27.141 | 47.201 | 8.986 | 2.386 | 1.324 | 0.284 | 0.155 | 0.063 |
| 1982 | 207.404 | 284.130 | 131.676 | 112.290 | 76.536 | 76.639 | 37.945 | 11.640 | 17.089 | 3.166 | 0.874 | 0.523 | 0.121 | 0.080 |
| 1983 | 209.808 | 169.808 | 232.626 | 104.869 | 75.856 | 42.234 | 35.952 | 15.469 | 3.899 | 5.417 | 1.105 | 0.347 | 0.228 | 0.063 |
| 1984 | 494.102 | 171.776 | 139.027 | 186.039 | 71.823 | 42.603 | 19.869 | 14.565 | 5.272 | 1.292 | 1.927 | 0.446 | 0.152 | 0.117 |
| 1985 | 390.766 | 404.537 | 140.638 | 109.517 | 124.653 | 40.300 | 20.524 | 8.302 | 5.353 | 2.052 | 0.533 | 0.890 | 0.217 | 0.085 |
| 1986 | 262.203 | 319.932 | 331.207 | 109.531 | 71.247 | 66.889 | 18.521 | 8.242 | 2.972 | 2.062 | 0.850 | 0.247 | 0.436 | 0.122 |
| 1987 | 133.093 | 214.674 | 261.938 | 255.023 | 69.046 | 34.796 | 26.832 | 6.668 | 2.616 | 1.029 | 0.800 | 0.371 | 0.116 | 0.236 |
| 1988 | 195.353 | 108.968 | 175.760 | 202.897 | 159.143 | 32.467 | 12.597 | 8.905 | 1.907 | 0.807 | 0.370 | 0.323 | 0.163 | 0.059 |
| 1989 | 159.674 | 159.942 | 89.215 | 137.298 | 128.377 | 77.221 | 12.016 | 4.117 | 2.440 | 0.541 | 0.266 | 0.133 | 0.127 | 0.074 |
| 1990 | 261.581 | 130.730 | 130.949 | 70.142 | 88.312 | 100.144 | 32.911 | 4.463 | 1.391 | 0.917 | 0.223 | 0.120 | 0.063 | 0.069 |
| 1991 | 202.807 | 214.164 | 107.033 | 101.969 | 44.722 | 45.142 | 42.298 | 12.288 | 1.564 | 0.549 | 0.390 | 0.103 | 0.059 | 0.035 |
| 1992 | 116.698 | 166.044 | 175.343 | 80.431 | 61.761 | 20.806 | 16.424 | 14.373 | 3.945 | 0.564 | 0.216 | 0.167 | 0.047 | 0.031 |
| 1993 | 226.858 | 95.545 | 135.945 | 129.645 | 47.837 | 27.798 | 7.150 | 5.367 | 4.371 | 1.360 | 0.216 | 0.090 | 0.075 | 0.024 |
| 1994 | 248.163 | 185.735 | 78.225 | 96.939 | 77.651 | 22.502 | 10.211 | 2.424 | 1.593 | 1.332 | 0.463 | 0.079 | 0.036 | 0.035 |
| 1995 | 133.267 | 203.179 | 152.067 | 58.639 | 62.370 | 43.331 | 10.847 | 4.275 | 0.936 | 0.655 | 0.569 | 0.212 | 0.038 | 0.020 |
| 1996 | 242.305 | 109.110 | 166.349 | 117.102 | 39.508 | 37.148 | 23.296 | 5.052 | 1.893 | 0.448 | 0.316 | 0.293 | 0.114 | 0.023 |
| 1997 | 106.298 | 198.382 | 89.332 | 131.373 | 81.718 | 24.410 | 20.181 | 10.971 | 2.239 | 0.889 | 0.211 | 0.159 | 0.156 | 0.067 |
| 1998 | 257.325 | 87.029 | 162.422 | 71.306 | 93.128 | 50.866 | 13.139 | 9.275 | 4.656 | 0.974 | 0.389 | 0.099 | 0.080 | 0.087 |
| 1999 | 243.543 | 210.680 | 71.254 | 129.222 | 50.124 | 54.943 | 24.854 | 5.589 | 3.545 | 1.757 | 0.376 | 0.160 | 0.045 | 0.041 |
| 2000 | 238.075 | 199.396 | 172.490 | 55.782 | 88.824 | 27.860 | 23.713 | 9.784 | 1.976 | 1.226 | 0.641 | 0.147 | 0.070 | 0.022 |
| 2001 | 267.383 | 194.919 | 163.252 | 133.003 | 38.175 | 49.533 | 12.405 | 9.388 | 3.437 | 0.672 | 0.434 | 0.241 | 0.062 | 0.033 |
| 2002 | 120.230 | 218.915 | 159.586 | 124.034 | 90.094 | 21.384 | 22.872 | 5.136 | 3.374 | 1.133 | 0.225 | 0.152 | 0.097 | 0.028 |
| 2003 | 231.193 | 98.436 | 179.232 | 124.690 | 86.284 | 52.743 | 10.802 | 10.349 | 2.074 | 1.255 | 0.414 | 0.086 | 0.066 | 0.046 |
| 2004 | 202.232 | 189.285 | 80.593 | 141.425 | 87.864 | 50.869 | 26.486 | 4.995 | 4.347 | 0.827 | 0.487 | 0.168 | 0.039 | 0.032 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 146.510 | 165.573 | 154.973 | 63.885 | 100.580 | 52.053 | 24.874 | 12.161 | 2.080 | 1.721 | 0.321 | 0.197 | 0.076 | 0.019 |
| 2006 | 198.692 | 119.952 | 135.560 | 122.911 | 46.117 | 61.663 | 26.582 | 11.798 | 5.233 | 0.857 | 0.697 | 0.134 | 0.091 | 0.038 |
| 2007 | 179.378 | 162.675 | 98.209 | 107.686 | 89.279 | 28.992 | 32.063 | 12.812 | 5.134 | 2.190 | 0.359 | 0.302 | 0.065 | 0.048 |
| 2008 | 194.898 | 146.863 | 133.187 | 78.267 | 79.141 | 58.116 | 16.232 | 16.184 | 5.771 | 2.191 | 0.935 | 0.158 | 0.149 | 0.034 |
| 2009 | 256.886 | 159.569 | 120.241 | 106.549 | 58.558 | 64.199 | 35.454 | 8.899 | 8.165 | 2.891 | 1.105 | 0.496 | 0.090 | 0.091 |
| 2010 | 271.227 | 210.320 | 130.644 | 95.566 | 79.327 | 39.797 | 38.840 | 19.449 | 4.510 | 4.169 | 1.515 | 0.614 | 0.296 | 0.058 |
| 2011 | 185.783 | 222.062 | 172.196 | 104.057 | 71.531 | 55.069 | 25.292 | 22.415 | 10.561 | 2.536 | 2.396 | 0.925 | 0.395 | 0.202 |
| 2012 | 275.001 | 152.106 | 181.809 | 137.269 | 77.937 | 49.787 | 35.465 | 14.977 | 12.613 | 6.228 | 1.527 | 1.534 | 0.620 | 0.280 |
| 2013 | 245.814 | 225.152 | 124.534 | 144.591 | 102.484 | 54.043 | 31.936 | 21.039 | 8.469 | 7.487 | 3.780 | 0.983 | 1.035 | 0.443 |
| 2014 | 175.214 | 201.255 | 184.339 | 97.953 | 107.046 | 70.308 | 34.306 | 18.852 | 11.771 | 4.958 | 4.467 | 2.383 | 0.653 | 0.730 |
| 2015 | 310.866 | 143.453 | 164.774 | 146.283 | 72.848 | 74.394 | 45.651 | 20.667 | 10.732 | 6.947 | 2.952 | 2.776 | 1.574 | 0.458 |
| 2016 | 310.293 | 254.516 | 117.450 | 130.346 | 108.457 | 50.631 | 48.840 | 27.929 | 11.966 | 6.455 | 4.187 | 1.848 | 1.863 | 1.126 |

Table 13. Icelandic cod in Division 5.a. Landings (thousand tonnes, average fishing mortality of age groups 5 to 10, recruitment to the fisheries at age 3 (millions), reference fishing biomass (B4+, thousand tonnes), spawning stock biomass (thousand tonnes) at spawning time and harvest ratio.

| Year | Yield | F5-10 | SSB | Reference biomass | Recruits | Harvest rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 545.250 | 0.288 | 953.278 | 2375.040 | 152.016 | 0.230 |
| 1956 | 486.909 | 0.288 | 806.298 | 2098.030 | 152.792 | 0.232 |
| 1957 | 455.182 | 0.309 | 785.078 | 1892.570 | 170.639 | 0.241 |
| 1958 | 517.359 | 0.352 | 883.799 | 1877.060 | 220.984 | 0.276 |
| 1959 | 459.081 | 0.321 | 859.929 | 1836.370 | 289.373 | 0.250 |
| 1960 | 470.121 | 0.370 | 712.439 | 1758.280 | 154.337 | 0.267 |
| 1961 | 377.291 | 0.354 | 469.374 | 1500.280 | 193.074 | 0.251 |
| 1962 | 388.985 | 0.382 | 571.614 | 1495.840 | 128.938 | 0.260 |
| 1963 | 408.800 | 0.457 | 510.186 | 1318.430 | 177.510 | 0.310 |
| 1964 | 437.012 | 0.547 | 453.246 | 1221.570 | 203.898 | 0.358 |
| 1965 | 387.106 | 0.575 | 318.588 | 1023.460 | 216.388 | 0.378 |
| 1966 | 353.357 | 0.588 | 277.969 | 1032.150 | 229.160 | 0.342 |
| 1967 | 335.721 | 0.560 | 257.009 | 1103.400 | 320.303 | 0.304 |
| 1968 | 381.770 | 0.720 | 221.959 | 1223.330 | 171.801 | 0.312 |
| 1969 | 403.205 | 0.557 | 314.034 | 1326.000 | 247.395 | 0.304 |
| 1970 | 475.077 | 0.610 | 331.527 | 1337.320 | 180.418 | 0.355 |
| 1971 | 444.248 | 0.683 | 242.925 | 1098.220 | 188.597 | 0.405 |
| 1972 | 395.166 | 0.692 | 222.252 | 997.449 | 139.216 | 0.396 |
| 1973 | 369.205 | 0.703 | 245.944 | 844.349 | 273.174 | 0.437 |
| 1974 | 368.133 | 0.762 | 187.669 | 919.100 | 178.982 | 0.401 |
| 1975 | 364.754 | 0.806 | 169.023 | 896.325 | 260.861 | 0.407 |
| 1976 | 346.253 | 0.742 | 139.357 | 956.674 | 367.683 | 0.362 |
| 1977 | 340.086 | 0.587 | 199.795 | 1291.280 | 143.320 | 0.263 |
| 1978 | 329.602 | 0.473 | 213.472 | 1299.590 | 227.589 | 0.254 |
| 1979 | 366.462 | 0.443 | 305.369 | 1398.860 | 243.438 | 0.262 |
| 1980 | 432.237 | 0.489 | 358.371 | 1491.730 | 140.028 | 0.290 |
| 1981 | 465.032 | 0.658 | 265.714 | 1243.880 | 140.305 | 0.374 |
| 1982 | 380.068 | 0.724 | 169.082 | 972.540 | 131.676 | 0.391 |
| 1983 | 298.049 | 0.708 | 132.004 | 793.209 | 232.626 | 0.376 |
| 1984 | 282.022 | 0.635 | 143.098 | 915.231 | 139.027 | 0.308 |
| 1985 | 323.428 | 0.663 | 165.594 | 928.906 | 140.638 | 0.348 |
| 1986 | 364.797 | 0.768 | 197.834 | 856.100 | 331.207 | 0.426 |
| 1987 | 389.915 | 0.854 | 152.472 | 1034.130 | 261.938 | 0.377 |
| 1988 | 377.554 | 0.883 | 169.571 | 1037.270 | 175.760 | 0.364 |
| 1989 | 363.125 | 0.710 | 174.449 | 1006.950 | 89.215 | 0.361 |
| 1990 | 335.316 | 0.692 | 215.606 | 843.547 | 130.949 | 0.398 |
| 1991 | 307.759 | 0.791 | 165.986 | 700.651 | 107.033 | 0.439 |
| 1992 | 264.834 | 0.833 | 152.494 | 553.175 | 175.343 | 0.479 |
| 1993 | 250.704 | 0.853 | 123.097 | 598.878 | 135.945 | 0.419 |
| 1994 | 178.138 | 0.612 | 158.854 | 580.018 | 78.225 | 0.307 |
| 1995 | 168.592 | 0.497 | 179.591 | 561.283 | 152.067 | 0.300 |
| 1996 | 180.701 | 0.495 | 161.850 | 676.221 | 166.349 | 0.267 |


| YEAR | YIELD | F5-10 | SSB | REFERENCE BIOMASS | RECRUITS | HARVEST RATE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 203.112 | 0.531 | 192.442 | 789.808 | 89.332 | 0.257 |
| 1998 | 243.987 | 0.631 | 204.859 | 729.154 | 162.422 | 0.335 |
| 1999 | 260.147 | 0.712 | 183.618 | 741.204 | 71.254 | 0.351 |
| 2000 | 235.092 | 0.714 | 173.878 | 602.061 | 172.490 | 0.390 |
| 2001 | 236.705 | 0.710 | 167.932 | 687.981 | 163.252 | 0.344 |
| 2002 | 209.537 | 0.619 | 203.796 | 731.859 | 159.586 | 0.286 |
| 2003 | 207.246 | 0.587 | 193.407 | 748.012 | 179.232 | 0.277 |
| 2004 | 228.337 | 0.594 | 202.944 | 809.602 | 80.593 | 0.282 |
| 2005 | 213.865 | 0.557 | 230.364 | 728.167 | 154.973 | 0.294 |
| 2006 | 197.247 | 0.537 | 225.566 | 704.547 | 135.560 | 0.280 |
| 2007 | 171.646 | 0.499 | 210.642 | 687.892 | 98.209 | 0.250 |
| 2008 | 147.668 | 0.389 | 273.051 | 710.846 | 133.187 | 0.208 |
| 2009 | 183.302 | 0.381 | 257.608 | 798.076 | 120.241 | 0.230 |
| 2010 | 170.009 | 0.318 | 295.656 | 858.715 | 130.644 | 0.198 |
| 2011 | 172.207 | 0.289 | 368.713 | 911.069 | 172.196 | 0.189 |
| 2012 | 196.177 | 0.287 | 408.345 | 1042.220 | 181.809 | 0.188 |
| 2013 | 223.594 | 0.299 | 447.024 | 1169.130 | 124.534 | 0.191 |
| 2014 | 221.990 | 0.285 | 417.504 | 1175.370 | 184.339 | 0.189 |
| 2015 | 230.225 | 0.273 | 533.186 | 1253.850 | 164.774 | 0.184 |
| 2016 | NA | NA | 464.020 | 1240.700 | 117.450 | NA |
| 2017 | NA | NA | NA | NA | 208.380 | NA |
| 2018 | NA | NA | NA | NA | 207.996 | NA |

Table 14. Icelandic cod in Division 5.a. Inputs in the deterministic predictions.

| Age | Parameter | 2016 | 2017 | 2018 | 2019 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | Catch weights | 1.401 | 1.401 | 1.401 | 1.401 |
| 4 | Catch weights | 1.815 | 1.815 | 1.815 | 1.815 |
| 5 | Catch weights | 2.265 | 2.265 | 2.265 | 2.265 |
| 6 | Catch weights | 3.500 | 3.500 | 3.500 | 3.500 |
| 7 | Catch weights | 4.153 | 4.153 | 4.153 | 4.153 |
| 8 | Catch weights | 5.338 | 5.338 | 5.338 | 5.338 |
| 9 | Catch weights | 6.664 | 6.664 | 6.664 | 6.664 |
| 10 | Catch weights | 8.603 | 8.603 | 8.603 | 8.603 |
| 11 | Catch weights | 9.688 | 9.688 | 9.688 | 9.688 |
| 12 | Catch weights | 11.212 | 11.212 | 11.212 | 11.212 |
| 13 | Catch weights | 11.334 | 11.334 | 11.334 | 11.334 |
| 14 | Catch weights | 10.356 | 10.356 | 10.356 | 10.356 |
| 3 | SSB weights | 0.973 | 0.973 | 0.973 | 0.973 |
| 4 | SSB weights | 2.223 | 2.223 | 2.223 | 2.223 |
| 5 | SSB weights | 3.035 | 3.035 | 3.035 | 3.035 |
| 6 | SSB weights | 4.198 | 4.198 | 4.198 | 4.198 |
| 7 | SSB weights | 4.610 | 4.610 | 4.610 | 4.610 |
| 8 | SSB weights | 6.018 | 6.018 | 6.018 | 6.018 |
| 9 | SSB weights | 7.389 | 7.389 | 7.389 | 7.389 |
| 10 | SSB weights | 9.731 | 9.731 | 9.731 | 9.731 |
| 11 | SSB weights | 9.688 | 9.688 | 9.688 | 9.688 |
| 12 | SSB weights | 11.212 | 11.212 | 11.212 | 11.212 |
| 13 | SSB weights | 11.334 | 11.334 | 11.334 | 11.334 |
| 14 | SSB weights | 10.356 | 10.356 | 10.356 | 10.356 |
| 3 | Maturity | 0.001 | 0.001 | 0.001 | 0.001 |
| 4 | Maturity | 0.009 | 0.009 | 0.009 | 0.009 |
| 5 | Maturity | 0.024 | 0.024 | 0.024 | 0.024 |
| 6 | Maturity | 0.288 | 0.288 | 0.288 | 0.288 |
| 7 | Maturity | 0.543 | 0.543 | 0.543 | 0.543 |
| 8 | Maturity | 0.733 | 0.733 | 0.733 | 0.733 |
| 9 | Maturity | 0.942 | 0.942 | 0.942 | 0.942 |
| 10 | Maturity | 0.987 | 0.987 | 0.987 | 0.987 |
| 11 | Maturity | 1.000 | 1.000 | 1.000 | 1.000 |
| 12 | Maturity | 1.000 | 1.000 | 1.000 | 1.000 |
| 13 | Maturity | 1.000 | 1.000 | 1.000 | 1.000 |
| 14 | Maturity | 1.000 | 1.000 | 1.000 | 1.000 |
| 3 | Selection | 0.123 | 0.123 | 0.123 | 0.123 |
| 4 | Selection | 0.345 | 0.345 | 0.345 | 0.345 |
| 5 | Selection | 0.589 | 0.589 | 0.589 | 0.589 |
| 6 | Selection | 0.825 | 0.825 | 0.825 | 0.825 |
| 7 | Selection | 1.080 | 1.080 | 1.080 | 1.080 |
| 8 | Selection | 1.273 | 1.273 | 1.273 | 1.273 |
| 9 | Selection | 1.134 | 1.134 | 1.134 | 1.134 |


| AGE | PARAMETER | 2016 | 2017 | 2018 | 2019 |
| :---: | :--- | :---: | :---: | :---: | :---: |
| 10 | Selection | 1.099 | 1.099 | 1.099 | 1.099 |
| 11 | Selection | 0.673 | 0.673 | 0.673 | 0.673 |
| 12 | Selection | 0.673 | 0.673 | 0.673 | 0.673 |
| 13 | Selection | 0.673 | 0.673 | 0.673 | 0.673 |
| 14 | Selection | 0.673 | 0.673 | 0.673 | 0.673 |
| 3 | Stock numbers | 117.450 | 208.380 | 207.996 | 0.000 |
| 4 | Stock numbers | 130.346 | NA | NA | NA |
| 5 | Stock numbers | 108.457 | NA | NA | NA |
| 6 | Stock numbers | 50.631 | NA | NA | NA |
| 7 | Stock numbers | 48.840 | NA | NA | NA |
| 8 | Stock numbers | 27.929 | NA | NA | NA |
| 9 | Stock numbers | 11.966 | NA | NA | NA |
| 10 | Stock numbers | 6.455 | NA | NA | NA |
| 11 | Stock numbers | 4.187 | NA | NA | NA |
| 12 | Stock numbers | 1.848 | NA | NA | NA |
| 13 | Stock numbers | 1.863 | NA | NA | NA |
| 14 | Stock numbers | 1.126 | NA | NA | NA |

Table 15. Icelandic cod in Division 5.a. Output of the deterministic predictions.

| Year | B4. | Fmult | Fbar | SSB | LANDINGS | 2018.B4. | 2018.SSB | SSB.Change | TAC.change |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | 1240.70 | NA | 0.30 | 464.020 | 246.000 | NA | NA |  |  |
| 2017 | 1191.34 | 0.00 | 0.00 | 548.116 | 0.000 | 1542.862 | 747.033 | 36\% | -100\% |
| NA | NA | 0.20 | 0.06 | 533.997 | 54.244 | 1481.375 | 686.418 | 29\% | -78\% |
| NA | NA | 0.23 | 0.07 | 531.682 | 63.011 | 1471.446 | 676.835 | 27\% | -74\% |
| NA | NA | 0.26 | 0.08 | 529.379 | 71.702 | 1461.605 | 667.394 | 26\% | -71\% |
| NA | NA | 0.30 | 0.09 | 527.086 | 80.318 | 1451.853 | 658.093 | 25\% | -67\% |
| NA | NA | 0.33 | 0.10 | 524.803 | 88.859 | 1442.187 | 648.930 | 24\% | -64\% |
| NA | NA | 0.36 | 0.11 | 522.532 | 97.327 | 1432.607 | 639.903 | 22\% | -60\% |
| NA | NA | 0.39 | 0.12 | 520.271 | 105.721 | 1423.112 | 631.010 | 21\% | -57\% |
| NA | NA | 0.43 | 0.13 | 518.021 | 114.043 | 1413.701 | 622.249 | 20\% | -54\% |
| NA | NA | 0.46 | 0.14 | 515.781 | 122.294 | 1404.374 | 613.617 | 19\% | -50\% |
| NA | NA | 0.49 | 0.15 | 513.552 | 130.473 | 1395.129 | 605.112 | 18\% | -47\% |
| NA | NA | 0.53 | 0.16 | 511.333 | 138.582 | 1385.967 | 596.734 | 17\% | -44\% |
| NA | NA | 0.56 | 0.17 | 509.125 | 146.622 | 1376.885 | 588.478 | 16\% | -40\% |
| NA | NA | 0.59 | 0.18 | 506.927 | 154.592 | 1367.883 | 580.345 | 14\% | -37\% |
| NA | NA | 0.63 | 0.19 | 504.739 | 162.495 | 1358.961 | 572.332 | 13\% | -34\% |
| NA | NA | 0.66 | 0.20 | 502.561 | 170.330 | 1350.117 | 564.436 | 12\% | -31\% |
| NA | NA | 0.69 | 0.21 | 500.394 | 178.098 | 1341.352 | 556.657 | 11\% | -28\% |
| NA | NA | 0.72 | 0.22 | 498.237 | 185.799 | 1332.663 | 548.993 | 10\% | -24\% |
| NA | NA | 0.76 | 0.23 | 496.090 | 193.436 | 1324.050 | 541.441 | 9\% | -21\% |
| NA | NA | 0.79 | 0.24 | 493.953 | 201.007 | 1315.513 | 534.000 | 8\% | -18\% |
| NA | NA | 0.82 | 0.25 | 491.826 | 208.514 | 1307.051 | 526.668 | 7\% | -15\% |
| NA | NA | 0.86 | 0.26 | 489.708 | 215.957 | 1298.662 | 519.444 | 6\% | -12\% |
| NA | NA | 0.89 | 0.27 | 487.601 | 223.337 | 1290.347 | 512.325 | 5\% | -9\% |
| NA | NA | 0.92 | 0.28 | 485.504 | 230.655 | 1282.105 | 505.311 | 4\% | -6\% |
| NA | NA | 0.95 | 0.29 | 483.416 | 237.911 | 1273.934 | 498.400 | 3\% | -3\% |
| NA | NA | 0.99 | 0.30 | 481.338 | 245.105 | 1265.835 | 491.590 | 2\% | -0\% |
| NA | NA | 1.02 | 0.31 | 479.270 | 252.239 | 1257.806 | 484.879 | 1\% | 3\% |
| NA | NA | 1.05 | 0.32 | 477.212 | 259.313 | 1249.846 | 478.266 | 0\% | 5\% |
| NA | NA | 1.09 | 0.33 | 475.163 | 266.328 | 1241.956 | 471.750 | -1\% | 8\% |
| NA | NA | 1.12 | 0.34 | 473.123 | 273.283 | 1234.134 | 465.329 | -2\% | 11\% |
| NA | NA | 1.15 | 0.35 | 471.093 | 280.180 | 1226.380 | 459.001 | -3\% | 14\% |
| NA | NA | 1.18 | 0.36 | 469.073 | 287.020 | 1218.693 | 452.766 | -3\% | 17\% |
| NA | NA | 1.22 | 0.37 | 467.062 | 293.802 | 1211.072 | 446.621 | -4\% | 19\% |
| NA | NA | 1.25 | 0.38 | 465.061 | 300.527 | 1203.517 | 440.566 | -5\% | 22\% |
| NA | NA | 1.28 | 0.39 | 463.068 | 307.197 | 1196.027 | 434.599 | -6\% | 25\% |
| NA | NA | 1.32 | 0.40 | 461.085 | 313.811 | 1188.601 | 428.719 | -7\% | 28\% |
| NA | NA | 1.35 | 0.41 | 459.111 | 320.370 | 1181.239 | 422.923 | -8\% | 30\% |
| NA | NA | 1.38 | 0.42 | 457.147 | 326.874 | 1173.940 | 417.212 | -9\% | 33\% |
| NA | NA | 1.41 | 0.43 | 455.191 | 333.324 | 1166.704 | 411.584 | -10\% | 35\% |
| NA | NA | 1.45 | 0.44 | 453.245 | 339.721 | 1159.530 | 406.037 | -10\% | 38\% |
| NA | NA | 1.48 | 0.45 | 451.308 | 346.065 | 1152.417 | 400.570 | -11\% | 41\% |
| NA | NA | 1.51 | 0.46 | 449.379 | 352.357 | 1145.365 | 395.183 | -12\% | 43\% |

9.3


Figure 1. Icelandic cod division 5.a. Comparisons of key stock assessment metrics using the old and the revised catch at age matrix.


Figure 2. Icelandic cod division 5.a. Estimated weight at age (numbers in panels indicate age classes) in the catches 1985-2016 expressed as deviation from the mean. Weights at age in 2016 are predicted from 2016 spring survey weights. Note that values that are equal to the mean are not visible in this type of a plot.


Figure 3. Icelandic cod division 5.a. Estimated weight at age (numbers in panel indicate age classes) Figure 3. Icelandic cod division 5.a. Estimated weight at age (numbers in panel indicate age classes) in the spring survey 1985-2016 (SMB) and fall survey 1996-2015 (SMH) expressed as proportional deviations from the mean. No fall survey was conducted in 2011. Note that values that are equal to the mean are not visible in this type of a plot.


Figure 4. Icelandic cod division 5.a. Abundance indices of cod in the groundfish survey in spring 1985-2016 (SMB red, longer time series) and fall 1996-2015 (SMH blue, shorter time series). Bottom left) Biomass index of 55 cm and larger, bottom right) Biomass index 80 cm and larger, top right) Abundance index of $<55 \mathrm{~cm}$, top left) Abundance index of $<18 \mathrm{~cm}$ fish. The shaded area and the vertical bar show 1 standard error of the estimate.


Figure 5. Icelandic cod division 5.a. Age based abundance indices of cod in the groundfish survey in spring 1985-2016(SMB) and fall 1996-2015 (SMH). The indices are standardized within each age group and within each survey.


Figure 6. Catch residuals (left), spring survey residuals (SMB, middle) and fall survey residuals (SMH, right) by year and age from the spaly ADCAM run. Note that values that are equal to the mean are not visible in this type of a plot and that no survey was carried out in the fall 2011.


Figure 7. Icelandic cod in division 5.a. Assessment summary based ADCAM tuned with the spring and the fall survey. The $x$-axis for the recruitment refer to the year class


Figure 8. Icelandic cod in division 5.a. Spawning stock biomass and corresponding recruitment at age 3. The numerical values refer to year class with the horizontal lines referring to geometric mean recruitment for year classes 1954-1984 (red line) and 1985-2016 (green line). Vertical lines refer to $B_{\text {lim }}\left(\boldsymbol{B}_{\text {loss }}\right.$, red) and $\boldsymbol{B}_{\text {trigger }}$ (green).


Figure 9. Icelandic cod in division 5.a. Empirical retrospective patterns from the 2004-2016 (this year's assessment, marked in red) assessments as summarized in ICES annual advisory sheet.


Figure 10. Icelandic cod in division 5.a. Distribution of realized harvest rate when TAC is based on the current catch rule ( $20 \%$ harvest rate and the catch stabilizer). The upper and lower $5 \%$ of the realized harvest rate is 0.15 and 0.27 . The distribution is based on the last 5 years in the simulations, the number of iterations being 1225 .


Figure 11. Icelandic cod in division 5.a. Median spawning stock biomass as a function of fishing mortality (age 5-10) (red line). The blue points indicate the historical estimates.


Figure 12. Icelandic cod in division 5.a. Mohn's rho statistics.

## 10 Icelandic haddock

The 2014 year class is estimated to be large, after 6 consecutive small year class from 2008-2013. The 2015 yearclass is expected to be close to geometric mean. The Current assessment shows some upward revision of the stock compared to last year's assessment. The main features are though the same that the fisheries are currently mostly based on relatively small year classes. It is first in 2018 that the 2014 year class affects the fisheries.

Growth in 2015 was above average since 1985 but less than predicted and the mean weight of young fish is above average while old fish are close to average. The assessment procedure was the same as last year (SPALY), an Adapt type model tuned with both the surveys.

There are differences in the perception of the state of stock in assessment based on either the spring or autumn survey with autumn survey indicating a larger stock. It has been like that since 2009. Different models using the same tuning data show similar results.

Advice is given according to the adopted Harvest Control Rule, and the advice for the fishing year 2016/2017 (September 1 ${ }^{\text {st }} 2016$-August 31 ${ }^{\text {st }} 2017$ ) is 34600 tonnes. The advice for the following fishing year is predicted to be approximately 43000 tonnes but increasing to more 50 thous. tonnes that when the 2014 year class is fully recruited.
No environmental drivers or ecosystem effects are known that can help in prediction of the development of the haddock stock. Some effect of the environment on the stock can though not be excluded.

### 10.1 Data

Landings of Icelandic haddock in 2015 are estimated to have been 39600 tonnes, see Figure 10.1.1 and Table 10.1.1. Of the landings, 38400 tonnes are caught by Iceland and 1200 tonnes by the Faeroese. The landings have decreased from 100 thous. tonnes between 2005-2008. The proportion of haddock caught by longliners was $44 \%$ in 2015, among the highest (Figure 10.1.2). On longer time-scale the share of longlines has increased in last 15 years, while the proportion of haddock caught in gillnets is now very small. Spatial distribution of the landings does not change very much from year-toyear but catches from the area north of Iceland have increased gradually over the last 10-15 years. (Figure 10.1.3 and 10.1.8).
Catch in numbers-at-age is shown in Table 10.1.2 and Figure 10.1.4. Age 8 accounted for $23 \%$ of the landings and age 9 and older for $10 \%$ while the average contribution of age 9 and older is $4.5 \%$. Age 8 the 2007 year class is the last above average year class in the fisheries so $67 \%$ of the catch is from the small year classes 2008-2013. The number of year classes contributing to the catches is unusually many, the result of low fishing mortality in recent years and the last large yearclass is 8 years old. The results for are close to expectation (Figure 10.1.5).
The index of total biomass from the groundfish surveys in March and October is shown in Figure 10.1.8. Both surveys show much increase between 2002 and 2005 but considerable decrease from 2007-2010. The difference in perception of the stock between the surveys is that the autumn survey shows less contrast between periods of large and small stock. In recent years the assessment has predicted reduced biomass while the
reality has been unchanged biomass in the March survey and some increase in the autumn survey, causing upwards revision of the stock in each assessment compared to earlier assessments.

Age disaggregated indices from the March survey are given in Table 10.1.3 and indices from the autumn survey in Table 10.1.4. Abundance of age groups 3-7 in the 2016 March survey is low while age 9 is among the highest indices observed (Figure 10.1.9). The index of age 12 and 13 (2003 cohort) is much higher than seen before (large part of $11+$ in the March survey), but that cohort will though not contribute much to the landings. Year classes 2008 and 2009 (age 8 and 7) are now close to average, mostly due to reduced fishing mortality in recent years but those year classes were originally small.

The survey results indicate that in recent decade higher and larger proportion of the haddock stock has gradually been inhabiting the waters north of Iceland (Figures 10.1.7 and 10.1.8.).

Mean weight at age in the catch is shown in Table 10.1.6 and Figure 10.1.10. Mean weight at age in the stock is given in Table 10.1.5 and Figure 10.1.9. Those data are obtained from the groundfish survey in March and are also used as mean weight at age in the spawning stock.

Both stock and catch weights have been increasing in recent years, after being very low when the stock was large between 2005-2009. Higher mean weight at age is most apparent for the younger haddock from the small cohorts (2008-2013), but mean weight of the old fish is now also average. Mean weight of the 2014 cohort is more than $20 \%$ lower than of recent small year classes but close to average for a large cohorts.

Prediction of growth is a source of uncertainty for this stock. (Figures 10.2.8, and 10.4.2). In recent year's growth has shown interannual variability without any pattern, indicating that short-term prediction should rather been based on average growth of last 2-3 years instead of only last year's growth. This approach might though have to be changed if stock size increases much so care should be exercised in carving any approach in stone

Maturity-at-age data are given in Table 10.1.7 and Figure 10.1.11. Those data are obtained from the groundfish survey in March. Maturity-at-age of the youngest age groups has been decreasing in recent years while mean weight at age has been increasing so maturity by size has been decreasing. The most likely explanation is large proportion of those age groups north of Iceland where proportion mature has always been low.

Catch per unit of effort data (figure 10.1.12) give somewhat different picture of the development of the stock than the surveys and assessment, much less increase after 2000 but much less decrease in recent years (figures 10.1.8vs.10.1.15). The interesting thing for the current assessment is the relatively high cpue, in recent years, confirming fishers's view that catching haddock in now very easy. The discrepancy observed between cpue and stock size has not been explained, but a number of plausible reasons mentioned.

- Area inhabited by the stock increased so the density in the traditional fishing area did not increase in relation to the stock size.
- When the stock was large slower growth lead to larger proportion of the stock below "fishable size" 45 cm limiting the areas where large haddock could be caught without too much bycatch of small haddock.
- The opposite is happening in recent years, faster growth and poor recruitment lead to the fisheries not limited by small haddock.
- Bycatch issues, but haddock is often caught as bycatch or one of the species in mixed fisheries where the goal is certain mixture of species.


### 10.2 Assessment.

From 200-2016 the final assessment was based on an Adapt type model calibrated with indices from both the groundfish surveys in March and October. Before that statistical catch-at-age model calibrated with indices from the March survey was used.

Assessment in recent years has shown some difference between different models, but more difference between different data sources i.e the March and the October surveys. From 2004-2008 models calibrated with the October survey indicated smaller stock. In the last five years things have changed and models calibrated with the October survey indicate a better state of the stock, the difference did not decrease with addition of the most recent data points i.e. October 2015 and March 2016. This behaviour is in line with what is seen in the surveys where the contrast in biomass is higher in the March survey (Figure 10.1.8).

The stock was benchmarked in February 2013, (WKROUND 2013) and the assessment procedure used since 2007 was recommended for few more years, if major problems do not show up (see stock annex).

The results of the assessment indicate that the stock decreased from 2008-2011 when large year classes disappeared from the stock and were replaced by smaller yea classes. (Figure 10.2.1) Since 2011 the rate of reduction has slowed down as fishing mortality has been low. The spawning stock has though decreased more than the reference biomass as proportion mature by age/size has been decreasing. Fishing mortality is now estimated to be low and should continue to be so if the adopted HCR will be followed and the stock size not overestimated. The current assessment does indicate the bottom has been reached and the stock size will increase in next years.

The main features of the current assessment are the same as in the assessments 2011 to 2015. The current assessment indicates larger stock than the 2015 assessment (Figures 10.2.7 and 10.2.8). Most of the difference is explained by higher than predicted num-bers-at-age (Figure 10.2.8). The tendency has been to underestimate recruitment and stock size in recent years.

Residuals from the assessment model are positive for the most recent October survey but close to zero for the most recent March survey. (Figures 10.2.2 and 10.2.3). The March surveys 2011-2015 are on the other hand below predictions. Similar thing seem to be happening in the fishery in 2012-2013 (Figure 10.1.15) so there are indication that the stock might be underestimated or availability of haddock is unusually high.

Standard errors in estimates of SSB in 2016 from the Adapt model are 9 thous. tons for the March survey and 16 thous. for the autumn survey. The difference between the stock biomass is 67 thous. tonnes ( 124 vs .57 thous. tonnes) that does not fit within the confidence intervals (less than $1 \%$ probability of 65 thous tonnes or more difference between autumn survey and March survey results). This is an indication that the estimated confidence intervals are too narrow. The same observation was made last 4 years. The spawning stock according to the model tuned with both the surveys is 77 thous. tonnes.

Plot of observed vs. predicted biomass from the surveys (figure 10.2.3) indicates that historically the autumn survey biomass has been closer to prediction than corresponding values from the March survey where the contrast in observed biomass is more than predicted from the assessment. When the stock was small in 2000 and 2001, the March survey indicated considerably smaller stock while the autumn survey values were reasonably correct and from 2003-2007 the March survey overestimated the stock.

There are indications that the autumn survey is a better predictor of haddock biomass than the March survey (Figure 10.2.3). The Adapt type model does though have some problems with using only the autumn survey and a separable model where selection is a function of stock weights (Björnsson 2013) gives biomass in 2016 as 108 thous. Tonnes when tuned only with the autumn survey. Also, including 10 shallow water stations added in 2008 might be questionable but excluding them gives $15 \%$ smaller stock when tuned with the autumn survey only. The conclusion is though that the current setting of tuning with both the surveys is a reasonable compromise, showing some underestimation in recent years. The assessment does also have many of the problems of low fishing mortality assessment i.e long periods of over/underestimation.

Figure 10.2 .5 shows the estimated "catchability" and CV as a function of age for the surveys, showing that estimated CV is lower in the autumn survey for ages $2-6$. Therefore, the autumn survey gets more weight for those age groups. The figure also indicates that estimated CV and "catchability" have not changed much for the March survey since 2008, but catchability of the autumn survey increased as has CV of the oldest age groups. This observation does partly have to do with the length of the series in.

To summarize there are indications from the autumn survey that the stock might be larger than predicted but from the March survey that it is smaller. cpue data, not used directly in the assessment support that the stock might be larger.

### 10.3 Reference points

In March 2013, ICES evaluated a proposed Harvest Control Rule for Icelandic haddock (Björnsson 2013) and the Icelandic government adopted it in April 2013. The Harvest control rule is

The annual total allowable catch (TAC) will be set by applying the following harvest control rule (HCR):

1. When spawing stock biomass in the year following the assessment year $\left(S S B_{y+1}\right)$ is equal to or greater than $S S B_{\text {trigger }}$ :

$$
\begin{aligned}
& \qquad T A C_{y / y+l}=\alpha B_{45+y+l} \\
& \text { 2. When } S S B_{y+1} \text { is below } S S B_{\text {trigger }} \text { : } \\
& T A C_{y y+1}=\alpha S S B_{y+1} / S S B_{\text {trigger }} \quad B_{45+, y+l}
\end{aligned}
$$

Where:
$y$ the assessment year,
$y / y_{+1} \quad$ the fishing year starting 1 September in year $y$ and ending 31 August in year $y+1$
$y_{-1} l y \quad$ the fishing year starting 1 September in year $y-1$ and ending 31 August in year $y$
$B_{45+, y+1}$ the reference biomass of 45 cm and larger haddock in the year following the assessment year and were $\alpha=0.40$ and $S S B_{\text {trigger }}=45000 \mathrm{t}$.
$B_{45+}$, is on the average close to the spawning stock, but is not affected by changes in proportion mature by size/age. Large variability of size at age (Figure 10.1.12) is the reason for basing reference biomass on size rather than age. Proportion of a cohort above $45 \mathrm{~cm}\left(\mathrm{~B}_{45+}\right)$ is calculated from stock weights by the green curve in Figure 10.4.3.

Blim for Icelandic haddock was defined by ICES in 2011 as 45000 tonnes or Bloss. From the simulations done to test the Harvest Control Rule $H_{m s y}$ the harvest ratio giving maximum yield was estimated as 0.52 and $H_{P A}$ harvest ratio giving 5\% probability of $\mathrm{SSB}<\boldsymbol{B}_{\lim }$ as 0.46 , compared to the target harvest rate of 0.4 . These numbers do though not have any meaning when the HCR has been adopted.

The reason for relatively low harvest ratio or 0.4 was to try to have low and stable harvest ratio, an approach considered appropriate strategy for stock with haddock like recruitment pattern, for example to avoid bycatch problems in mixed fisheries in periods of poor recruitment.

When the HCR was evaluated in 2013 the understanding was the only reference point needed was Blim that was to be avoided by more than $95 \%$ probability according to the management plan.

At this meeting classification reference points $B_{p a}, F_{\text {lim }}$ and $F_{p a}$ were required. $B_{p a}$ was supposed to be defined in such a way that if the stock was estimated at Bpa the probability of being below Blim was less than $5 \%$. Estimated CV of spawning stock at the beginning of the assessment year assessment year is $16 \%$ (Björnsson 2013) so $B_{p a}=$ $B_{\text {lim }} \times e^{1.645 \sigma}=45 \times e^{1.645 \times 0.16}=59$ thous. Tonnes.

To get values of $F_{l i m}$ and $F_{p a}$ or really $H_{l i m}$ and $H_{p a}$ the model used for HCR evaluation was run with 3 changes in settings.

1. No autocorrelation of recruitment.
2. No assessment error.
3. No trigger.

The model is still stochastic, taking into account variability of recruitment as well as uncertainty in the parameters of the stock recruitment function and other parameters. What was used as basis for $H_{\text {lim }}$ was the harvest rate where the average of the spawning stock was equal to $B_{\text {lim }}$ The results of the model runs are show that $H_{\text {lim }}=0.9$ and $H_{p a}=B_{l i m} \times e^{-1.645 \sigma}=0.7$. (Figure 10.3.1). According to the HCR simulations the fifth percentile of SSB would hardly be positive if fishing at $H_{l i m}$ or even at $H_{p a}$. If a completely deterministic model was run the value of $\mathrm{H}_{\text {lim }}$ obtained would be even higher. The purpose of these values is hard to understand, but they have at least been provided.

According to the simulations done in 2013 the probability of $\mathrm{SSB}<\mathrm{B}_{\lim }$ was 0.07 . Going back to the same results and the recently "invented" $\mathrm{B}_{\mathrm{pa}} \mathrm{SSB}<\mathrm{B}_{\mathrm{pa}}=0.07$ but the probability of estimated $\mathrm{SSB}<\mathrm{B}_{\mathrm{pa}}$ is higher or $0.19 \%$, i.e the probability of red light was $17 \%$. As the HCR takes into account stochasticity in recruitment analysis based on fifth percentile of SSB lead to low probability of $H>H_{p a}$ (figure 10.3.2)

Those analysis show that the recently invented reference point should not be in the way of sound management of this stock. The reason is relatively large reduction in fishing pressure from historic values and no catch stabilizer. For stocks where either or both of those condition is not met (stocks that have never been overexploited) there might be a problem.

### 10.4 Short-term forecast

Prediction of weight at age in the stock, weight at age in the catches, maturity-at-age and selection has been similar since 2006 (WD \#19 in 2006). The procedure is described in the advice part of the report of ADGISHA (Björnsson 2013) and also in the stock annex. The procedure was changed last year so instead of taking only last years value, average of last 2 values is used.

To summarize, TAC for the fishing year 2016/2017 is a function of the biomass of 45 cm and larger haddock and the spawning stock at the beginning of 2016. To be able to predict the stock size in 2017, catch 2016, mean weight at age in the catch 2016, selection at age in the catch 2016, stock weights in 2017 and maturity-at-age in 2017 must be predicted. The prediction of these values is described in Björnsson (2013) and the stock annex, but to summarize, catch in the assessment year (2016) is the TAC left in the current fishing year at the beginning of the assessment year plus $1 / 3$ of the predicted TAC next fishing year. The TAC for the fishing year 2015/2016 was 36400 tonnes. The landings in September-December 2014 were 14400 tonnes or $39 \%$ of the TAC. The average contribution of the first 4 month of the Fishing year is on the other hand around $33 \%$. Landings for the fishing year 2014/2015 are now estimated to be 36600 tonnes while the TAC issued was 30400 . tonnes. Looking at the rate of landings (Figure 10.4.1) they indicate that the TAC for the current fishing year could be exceeded. .

In the Icelandic fishery management system certain relatively small transfer is allowed between species, to increase flexibility in mixed fisheries. Currently net transfer is towards haddock, probably because haddock is easy to catch, as demonstrated by high cpue in 2015. The haddock quota does also seem to be limiting in some mixed fisheries. The reason that haddock has been underestimated in last years could also contribute to transfer towards haddock. Looking over longer period quota transfer towards/from haddock has on the average been close to zero. In predictions for current fishing year 1000 ton transfer towards haddock is assumed.

On January $1^{\text {st }} 2016,22$ thous. tonnes of quota were left. To this are added $1 / 3^{\text {rd }}$ of next year's TAC ( 11500000 tonnes). This leads to 33.500 tonnes catch in the calendar year 2015.

Mean weight and maturity-at-age in 2016 are available and are used to predict catch weights and selection at age (Figure 10.4.2). Growth in 2016 is predicted by the equation

$$
\log \frac{W_{a+1, t+1}}{W_{a, t}}=\alpha+\beta \log W_{a, t}+\delta_{y e a r}
$$

Where according to the stock annex the factor $\delta_{\text {year }}$ for the assessment year (figure
$\square$ 10.4.2) is the average of $\delta_{\text {year }}$ of the growth in the 2 preceding years. Growth has been high but somewhat variable in recent years but was much less in when the stock was larger (figure 10.4.2).

Maturity, selection, catch weights at age and proportion of the biomass above 45 cm are then predicted from stock weights 2017. When those values have been estimated the prediction is done by the same model as used in the assessment.

The model works iteratively as the estimated TAC for the fishing year 2016/2017 has some effect of the biomass at the beginning of 2017, which the TAC is based on. Advice
for the following fishing year (2017/2018) is predicted to be approximately 42000 tonnes but increasing after that when the 2014 year is fully recruited.
Results of the short-term prediction are shown in figure 10.2.1 assuming that the harvest control rule is followed. TAC for the fishing year 2016/2017 will be 34600 tons. Short-term prognosis based on the traditional ICES approach are shown in table 10.4.1

### 10.5 References

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Table 10.1.1 Haddock in Division Va Landings by nation.

| COUNTRY | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 1010 | 1144 | 673 | 377 | 268 | 359 | 391 | 257 |
| Faroe Islands | 2161 | 2029 | 1839 | 1982 | 1783 | 707 | 987 | 1289 |
| Iceland | 52152 | 47916 | 61033 | 67038 | 63889 | 47216 | 49553 | 47317 |
| Norway | 11 | 23 | 15 | 28 | 3 | 3 | + |  |
| UK |  |  |  |  |  |  |  |  |
| Total | 55334 | 51112 | 63560 | 69425 | 65943 | 48285 | 50933 | 48863 |

HADDOCK Va

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 238 | 352 | 483 | 595 | 485 | 361 | 458 | 248 |
| Faroe Islands | 1043 | 797 | 606 | 603 | 773 | 757 | 754 | 911 |
| Iceland | 39479 | 53085 | 61792 | 66004 | 53516 | 46098 | 46932 | 58408 |
| Norway | 1 | + |  |  |  |  |  | 1 |
| Total | 40761 | 54234 | 62881 | 67202 | 53774 | 47216 | 48144 | 59567 |

HADDOCK Va

| Country | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  |  |  |  |  |  |  |  |
| Faroe Islands | 758 | 664 | 340 | 639 | 624 | 968 | 609 | 878 |
| Iceland | 60061 | 56223 | 43245 | 40795 | 44557 | 41199 | 39038 | 49591 |
| Norway | + | 4 |  |  |  |  |  |  |
| Total | 60819 | 56891 | 43585 | 41434 | 45481 | 42167 | 39647 | 50469 |


| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  |  |  |  |  |  |  |  |  |
| Faroe <br> Islands | 833 | 1035 | 1372 | 1499 | 1780 | 828 | 625 | 311 | 207 |
| Iceland | 59970 | 83791 | 95859 | 96115 | 108175 | 101651 | 81418 | 63868 | 49231 |
| Norway | 30 | 9 |  |  | 11 | 11 |  |  |  |
| Total | 60884 | 84835 | 97231 | 97614 | 109966 | 102490 | 82043 | 64179 | 49437 |


| Country | 2012 | 2013 | 2014 | 2015 |
| :--- | :---: | :---: | :---: | :---: |
| Belgium |  |  |  |  |
| Faroe <br> Islands | 303 | 600 | 800 | 1259 |
| Iceland | 45888 | 43500 | 33100 | 38391 |
| Norway |  |  |  |  |
| Total | 46191 | 44100 | 33900 | 39650 |

Table 10.1.2 Haddock in division Va. Catch in number by year and age.

| Year/ <br> Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 149 | 1908 | 3762 | 6057 | 9022 | 1743 | 438 | 56 | 112 |
| 1980 | 595 | 1385 | 11481 | 4298 | 3798 | 3732 | 544 | 91 | 37 |
| 1981 | 10 | 514 | 4911 | 16900 | 5999 | 2825 | 1803 | 168 | 57 |
| 1982 | 107 | 245 | 3149 | 10851 | 14049 | 2068 | 1000 | 725 | 201 |
| 1983 | 34 | 1010 | 1589 | 4596 | 9850 | 8839 | 766 | 207 | 280 |
| 1984 | 241 | 1069 | 4946 | 1341 | 4772 | 3742 | 4076 | 238 | 80 |
| 1985 | 1320 | 1728 | 4562 | 6796 | 855 | 1682 | 1914 | 1903 | 296 |
| 1986 | 1012 | 4223 | 4068 | 4686 | 5139 | 494 | 796 | 897 | 400 |
| 1987 | 1939 | 8308 | 6965 | 2728 | 2042 | 1094 | 132 | 165 | 339 |
| 1988 | 237 | 9831 | 15164 | 5824 | 1304 | 1084 | 609 | 66 | 213 |
| 1989 | 188 | 2474 | 22560 | 9571 | 3196 | 513 | 556 | 144 | 141 |
| 1990 | 1857 | 2415 | 8628 | 23611 | 6331 | 816 | 150 | 67 | 74 |
| 1991 | 8617 | 2145 | 5397 | 7342 | 14103 | 2648 | 338 | 40 | 27 |
| 1992 | 5405 | 10693 | 5721 | 4610 | 3691 | 5209 | 999 | 120 | 16 |
| 1993 | 769 | 12333 | 12815 | 2968 | 1722 | 1425 | 2239 | 343 | 38 |
| 1994 | 3198 | 3343 | 28258 | 10682 | 1469 | 726 | 358 | 647 | 108 |
| 1995 | 4015 | 7323 | 5744 | 23927 | 5769 | 615 | 290 | 187 | 331 |
| 1996 | 3090 | 10552 | 7639 | 4468 | 12896 | 2346 | 208 | 79 | 125 |
| 1997 | 1364 | 3939 | 10915 | 4895 | 2610 | 5035 | 719 | 64 | 69 |
| 1998 | 279 | 8257 | 5667 | 7856 | 2418 | 1422 | 1897 | 261 | 45 |
| 1999 | 1434 | 1550 | 17243 | 4516 | 4837 | 915 | 620 | 481 | 64 |
| 2000 | 2659 | 6317 | 2352 | 13615 | 1945 | 1706 | 324 | 222 | 192 |
| 2001 | 2515 | 11098 | 6954 | 1446 | 6262 | 675 | 478 | 105 | 94 |
| 2002 | 1082 | 10434 | 15998 | 5099 | 1131 | 3149 | 262 | 169 | 100 |
| 2003 | 401 | 6352 | 16265 | 12548 | 2968 | 748 | 1236 | 91 | 70 |
| 2004 | 1597 | 4063 | 17652 | 19358 | 8871 | 1940 | 471 | 489 | 155 |
| 2005 | 2405 | 9450 | 6929 | 25421 | 13778 | 4584 | 809 | 251 | 237 |
| 2006 | 241 | 10038 | 21246 | 6646 | 18840 | 7600 | 2180 | 323 | 202 |
| 2007 | 782 | 3884 | 42224 | 22239 | 3354 | 9952 | 2740 | 519 | 181 |
| 2008 | 2316 | 4508 | 9706 | 53022 | 11014 | 1717 | 3033 | 815 | 192 |
| 2009 | 1066 | 3185 | 4886 | 8892 | 35011 | 5733 | 726 | 1381 | 509 |
| 2010 | 121 | 6032 | 7061 | 4806 | 6766 | 17503 | 1874 | 354 | 528 |
| 2011 | 253 | 1584 | 11797 | 5080 | 2853 | 3983 | 6220 | 494 | 183 |
| 2012 | 196 | 1322 | 3421 | 13107 | 2223 | 1231 | 2480 | 2662 | 370 |
| 2013 | 250 | 1042 | 2865 | 4008 | 9222 | 1206 | 668 | 1248 | 1599 |
| 2014 | 238 | 1478 | 1751 | 2725 | 2737 | 4742 | 447 | 387 | 1403 |

Table 10.1.3 Icelandic haddock. Age disaggregated survey indices from the groundfish survey in March.

| Year/ <br> Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | $11+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 28.14 | 32.68 | 18.33 | 23.58 | 26.39 | 3.7 | 10.86 | 4.8 | 5.54 | 0.49 | 0.19 |
| 1986 | 123.87 | 108.48 | 58.97 | 12.79 | 16.31 | 13.12 | 0.97 | 2.71 | 1.22 | 2.25 | 0.19 |
| 1987 | 21.82 | 338.29 | 147.5 | 44.15 | 7.68 | 7.47 | 4.72 | 0.39 | 0.61 | 0.44 | 0.86 |
| 1988 | 15.77 | 40.73 | 184.79 | 88.87 | 22.86 | 1.34 | 2.18 | 1.76 | 0.16 | 0.22 | 0.31 |
| 1989 | 10.58 | 23.33 | 41.16 | 146.61 | 45.09 | 12.88 | 0.79 | 0.81 | 0.41 | 0.28 | 0.23 |
| 1990 | 70.48 | 31.8 | 26.73 | 38.84 | 92.82 | 30.89 | 3.44 | 0.88 | 0.23 | 0 | 0.02 |
| 1991 | 89.73 | 145.95 | 41.43 | 17.73 | 20.19 | 32.85 | 7.63 | 0.3 | 0.1 | 0.08 | 0.08 |
| 1992 | 18.15 | 211.43 | 137.77 | 35.38 | 16.91 | 13.77 | 16.32 | 2.22 | 0.18 | 0.07 | 0 |
| 1993 | 29.99 | 37.8 | 244.96 | 87.19 | 11.23 | 3.85 | 1.66 | 4.46 | 0.88 | 0 | 0 |
| 1994 | 58.54 | 61.34 | 39.83 | 142.35 | 42.18 | 6.9 | 2.87 | 1.42 | 4.44 | 0.17 | 0 |
| 1995 | 35.89 | 82.47 | 47.03 | 19.75 | 69.52 | 7.66 | 1.31 | 0.11 | 0.34 | 0 | 0 |
| 1996 | 95.25 | 66.21 | 119.86 | 36.78 | 19.58 | 40.63 | 5.78 | 0.59 | 0.13 | 0.12 | 0.15 |
| 1997 | 8.6 | 119.35 | 50.81 | 53.33 | 10.88 | 7.37 | 10.9 | 1.35 | 0.07 | 0.03 | 0.13 |
| 1998 | 23.08 | 18 | 107.93 | 28.23 | 23.49 | 4.9 | 3.54 | 4.56 | 0.33 | 0 | 0 |
| 1999 | 80.73 | 85.46 | 25.53 | 98.73 | 12.99 | 9.85 | 1.42 | 1.77 | 1.03 | 0.09 | 0 |
| 2000 | 60.58 | 90.07 | 44.63 | 8.45 | 25.22 | 3.14 | 1.59 | 0.4 | 0.15 | 0.52 | 0.04 |
| 2001 | 81.27 | 147.71 | 115.4 | 22.15 | 4.09 | 10.63 | 0.93 | 0.57 | 0 | 0.1 | 0 |
| 2002 | 20.75 | 298.67 | 200.74 | 112.49 | 23.24 | 3.51 | 7.49 | 0.31 | 0.3 | 0.08 | 0.15 |
| 2003 | 111.59 | 97.54 | 282.28 | 244.81 | 113.45 | 18 | 2.55 | 4.48 | 0.48 | 0.82 | 0.15 |
| 2004 | 325.9 | 291.65 | 70.75 | 208.74 | 109.33 | 33.96 | 6.79 | 1.24 | 0.82 | 0 | 0.31 |
| 2005 | 57.96 | 698.48 | 289.43 | 44.58 | 157.2 | 57.52 | 15.72 | 3.35 | 0.32 | 0.25 | 0.02 |
| 2006 | 39.29 | 88.69 | 575.93 | 179.11 | 19.13 | 62.94 | 16.43 | 6.74 | 0.7 | 0.29 | 0 |
| 2007 | 34 | 65.6 | 88.63 | 436.41 | 85.68 | 7.9 | 21.6 | 4.74 | 2.15 | 0.07 | 0 |
| 2008 | 88.53 | 68.05 | 71.7 | 75.57 | 222.79 | 29.99 | 3.53 | 7.47 | 1.64 | 0.27 | 0.03 |
| 2009 | 10.46 | 111.21 | 53.82 | 41.48 | 41.91 | 105.64 | 12.94 | 2.23 | 3.11 | 0.44 | 0.23 |
| 2010 | 15.15 | 27.71 | 138.2 | 29.95 | 18.28 | 20.59 | 31.59 | 2.92 | 0.46 | 0.69 | 0.2 |
| 2011 | 8.79 | 27.65 | 24.75 | 77.43 | 14.03 | 5.9 | 9.4 | 14.89 | 1.22 | 0.31 | 0.3 |
| 2012 | 12.47 | 14.9 | 31.27 | 27.22 | 58.3 | 5.23 | 2.92 | 5.3 | 6.87 | 0.8 | 0.49 |
| 2013 | 13.91 | 23.32 | 19.72 | 22.9 | 22.51 | 41.93 | 4.78 | 2.52 | 3.83 | 4.52 | 1.02 |
| 2014 | 14.01 | 24.78 | 30.27 | 17.74 | 16.44 | 14.79 | 16.44 | 1.33 | 1.05 | 1.68 | 1.63 |
| 2015 | 62.58 | 19.59 | 26.56 | 34.23 | 12.58 | 11.18 | 9.63 | 9.96 | 1.14 | 0.56 | 2.29 |
| 2016 | 30.02 | 163.8 | 4.08 | 22.2 | 22.26 | 7.17 | 7.27 | 5.05 | 4.2 | 0.93 | 1.79 |

Table 10.1.4 Icelandic haddock. Age disaggregated survey indices from the groundfish survey in October

| YEAR/AGE | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 16 | 458 | 108 | 83.9 | 18 | 7.6 | 17.6 | 1.5 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 52 | 32 | 210 | 53.5 | 37.6 | 6.8 | 5.6 | 5.8 | 0.3 | 0 | 0 | 0 | 0 |
| 1998 | 208 | 81 | 32 | 131.1 | 19.3 | 15.2 | 5 | 5.2 | 1.8 | 0 | 0 | 0.07 | 0 |
| 1999 | 174 | 396 | 66 | 28.3 | 95.7 | 11.6 | 10.1 | 0.5 | 2.1 | 0.29 | 0 | 0 | 0 |
| 2000 | 54 | 161 | 259 | 45.8 | 8.1 | 28.3 | 1.9 | 3.2 | 0.1 | 0.27 | 0.58 | 0 | 0 |
| 2001 | 46 | 382 | 277 | 172.1 | 34.9 | 3.9 | 13.9 | 0.7 | 0.9 | 0 | 0.21 | 0 | 0 |
| 2002 | 148 | 80 | 239 | 189.7 | 94.1 | 18.4 | 2.8 | 2.1 | 1 | 0.04 | 0 | 0 | 0 |
| 2003 | 315 | 344 | 145 | 247.6 | 164.9 | 54.5 | 8.9 | 2.4 | 0.6 | 0 | 0.04 | 0 | 0 |
| 2004 | 187 | 709 | 344 | 50 | 156.1 | 68.1 | 16.2 | 3.9 | 0.8 | 0.49 | 0 | 0 | 0 |
| 2005 | 90 | 73 | 552 | 178.9 | 26.4 | 93.6 | 25.5 | 9.7 | 1.8 | 0 | 0.12 | 0 | 0 |
| 2006 | 84 | 124 | 116 | 500.6 | 105.7 | 13.4 | 39.4 | 9.4 | 3.9 | 1.5 | 0 | 0 | 0 |
| 2007 | 233 | 97 | 78 | 89.2 | 328 | 56.8 | 7.9 | 12 | 3.6 | 0.54 | 0.19 | 0 | 0.09 |
| 2008 | 95 | 201 | 93 | 67.1 | 85.7 | 193.6 | 16.3 | 2.8 | 3.3 | 0.21 | 0.07 | 0 | 0 |
| 2009 | 51 | 47 | 268 | 67.2 | 30.4 | 47.5 | 94.2 | 9.2 | 1.4 | 2.09 | 0.05 | 0.36 | 0 |
| 2010 | 36 | 42 | 56 | 141.6 | 30 | 14.1 | 23.2 | 36.3 | 4.6 | 0.85 | 0.95 | 0.15 | 0 |
| 2012 | 26 | 53 | 29 | 33.7 | 37.1 | 69.2 | 9.1 | 3.5 | 9.6 | 10.09 | 0.97 | 0.18 | 0.5 |
| 2013 | 27 | 90 | 127 | 36.5 | 37.8 | 38.7 | 44.2 | 6.2 | 2.3 | 5.69 | 4.14 | 0.69 | 0 |
| 2014 | 248 | 34 | 41 | 65.5 | 23.4 | 26.4 | 23.8 | 25.8 | 2.2 | 1.46 | 2.94 | 1.44 | 0.54 |
| 2015 | 132 | 204 | 36 | 38.7 | 47.7 | 15.1 | 18 | 10.3 | 12 | 2.26 | 1.36 | 0.54 | 1.35 |

Table 10.1.5 Haddock in division Va Weight at age in the stock. Predicted values are shaded

| Year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 4000 |
| 1980 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 4615 |
| 1981 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 4898 |
| 1982 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 3952 |
| 1983 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 4463 |
| 1984 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 3941 |
| 1985 | 36 | 244 | 568 | 1187 | 1673 | 2371 | 2766 | 3197 | 3331 | 4564 |
| 1986 | 35 | 239 | 671 | 1134 | 1943 | 2399 | 3190 | 3293 | 3728 | 4436 |
| 1987 | 31 | 162 | 550 | 1216 | 1825 | 2605 | 3030 | 3642 | 3837 | 3653 |
| 1988 | 37 | 176 | 457 | 974 | 1830 | 2695 | 3102 | 3481 | 3318 | 4169 |
| 1989 | 26 | 182 | 441 | 887 | 1510 | 2380 | 3009 | 3499 | 3195 | 5039 |
| 1990 | 29 | 184 | 457 | 840 | 1234 | 1965 | 2675 | 3052 | 3267 | 4115 |
| 1991 | 31 | 176 | 501 | 1003 | 1406 | 1884 | 2496 | 3755 | 3653 | 5243 |
| 1992 | 28 | 157 | 503 | 894 | 1365 | 1891 | 2325 | 2936 | 3682 | 4674 |
| 1993 | 41 | 168 | 384 | 878 | 1492 | 1785 | 2562 | 2573 | 3266 | 4047 |
| 1994 | 33 | 181 | 392 | 680 | 1235 | 1766 | 1717 | 2977 | 2131 | 3154 |
| 1995 | 37 | 167 | 440 | 755 | 1065 | 1857 | 2689 | 5377 | 1306 | 3119 |
| 1996 | 41 | 174 | 453 | 813 | 1076 | 1477 | 2171 | 2426 | 4847 | 3686 |
| 1997 | 50 | 174 | 424 | 817 | 1221 | 1425 | 1915 | 2390 | 3692 | 3508 |
| 1998 | 41 | 203 | 415 | 753 | 1241 | 1747 | 1996 | 2342 | 3076 | 3275 |
| 1999 | 33 | 206 | 480 | 715 | 1189 | 1956 | 2366 | 2782 | 2922 | 3534 |
| 2000 | 29 | 179 | 552 | 889 | 1159 | 1767 | 2612 | 2917 | 3132 | 3734 |
| 2001 | 36 | 190 | 490 | 1056 | 1437 | 1509 | 2169 | 2765 | 3300 | 4715 |
| 2002 | 67 | 172 | 475 | 889 | 1460 | 1949 | 2137 | 1990 | 3709 | 4078 |
| 2003 | 40 | 230 | 412 | 801 | 1268 | 1873 | 3139 | 2343 | 3301 | 3289 |
| 2004 | 34 | 176 | 556 | 807 | 1282 | 1690 | 2454 | 3236 | 2942 | 3957 |
| 2005 | 40 | 153 | 448 | 920 | 1188 | 1564 | 2128 | 2808 | 2550 | 2755 |
| 2006 | 33 | 127 | 333 | 736 | 1145 | 1512 | 1944 | 2232 | 3272 | 3617 |
| 2007 | 48 | 170 | 350 | 615 | 1053 | 1514 | 1786 | 2073 | 2198 | 2408 |
| 2008 | 27 | 179 | 382 | 595 | 868 | 1295 | 1828 | 2201 | 2340 | 2568 |
| 2009 | 29 | 139 | 442 | 687 | 882 | 1141 | 1495 | 1920 | 2574 | 3070 |
| 2010 | 32 | 150 | 392 | 773 | 942 | 1190 | 1468 | 1829 | 2086 | 2730 |
| 2011 | 35 | 175 | 442 | 757 | 1129 | 1304 | 1583 | 1865 | 2107 | 3094 |
| 2012 | 28 | 202 | 482 | 801 | 1145 | 1480 | 1909 | 2072 | 2353 | 2350 |
| 2013 | 33 | 201 | 589 | 967 | 1312 | 1710 | 1999 | 2265 | 2764 | 2709 |
| 2014 | 36 | 222 | 570 | 1005 | 1372 | 1751 | 2141 | 2298 | 2653 | 3104 |
| 2015 | 32 | 255 | 614 | 1073 | 1637 | 1926 | 2452 | 2774 | 3170 | 3173 |
| 2016 | 29 | 162 | 642 | 1099 | 1564 | 2094 | 2296 | 3068 | 3481 | 3248 |
| 2017 | 29 | 181 | 454 | 1174 | 1701 | 2170 | 2654 | 2828 | 3454 | 3769 |
| 2018 | 29 | 187 | 489 | 924 | 1780 | 2299 | 2720 | 3125 | 3265 | 3749 |


| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 620 | 960 | 1410 | 2030 | 2910 | 3800 | 4560 | 4720 | 5956 |
| 1980 | 837 | 831 | 1306 | 2207 | 2738 | 3188 | 3843 | 4506 | 4983 |
| 1981 | 584 | 693 | 1081 | 1656 | 2283 | 3214 | 3409 | 4046 | 5261 |
| 1982 | 289 | 959 | 1455 | 1674 | 2351 | 3031 | 3481 | 3874 | 4123 |
| 1983 | 320 | 1006 | 1496 | 1921 | 2371 | 2873 | 3678 | 4265 | 4502 |
| 1984 | 691 | 1007 | 1544 | 2120 | 2514 | 3027 | 2940 | 3906 | 4033 |
| 1985 | 652 | 1125 | 1811 | 2260 | 2924 | 3547 | 3733 | 4039 | 4659 |
| 1986 | 336 | 1227 | 1780 | 2431 | 2771 | 3689 | 3820 | 4258 | 4456 |
| 1987 | 452 | 1064 | 1692 | 2408 | 3000 | 3565 | 4215 | 4502 | 4025 |
| 1988 | 362 | 780 | 1474 | 2217 | 2931 | 3529 | 3781 | 4467 | 4418 |
| 1989 | 323 | 857 | 1185 | 1996 | 2893 | 4066 | 3866 | 4734 | 4990 |
| 1990 | 269 | 700 | 1054 | 1562 | 2364 | 3414 | 4134 | 4946 | 4451 |
| 1991 | 288 | 699 | 979 | 1412 | 1887 | 2674 | 3135 | 4341 | 4957 |
| 1992 | 313 | 806 | 1167 | 1524 | 1950 | 2357 | 3075 | 4053 | 4703 |
| 1993 | 303 | 705 | 1333 | 1875 | 2386 | 2996 | 3059 | 3363 | 4409 |
| 1994 | 337 | 668 | 1019 | 1717 | 2391 | 2717 | 3280 | 3156 | 3278 |
| 1995 | 351 | 746 | 1096 | 1318 | 2044 | 2893 | 3049 | 3675 | 3137 |
| 1996 | 311 | 787 | 1187 | 1560 | 1849 | 2670 | 3510 | 3567 | 3731 |
| 1997 | 379 | 764 | 1163 | 1649 | 1943 | 2342 | 3020 | 3337 | 3236 |
| 1998 | 445 | 724 | 1147 | 1683 | 2250 | 2475 | 2834 | 3333 | 3596 |
| 1999 | 555 | 908 | 1101 | 1658 | 2216 | 2659 | 2928 | 3209 | 3513 |
| 2000 | 495 | 978 | 1333 | 1481 | 2119 | 2696 | 3307 | 3597 | 3757 |
| 2001 | 541 | 945 | 1456 | 1731 | 1832 | 2243 | 3020 | 3328 | 4236 |
| 2002 | 564 | 928 | 1253 | 1737 | 2219 | 2230 | 2911 | 3365 | 4387 |
| 2003 | 498 | 922 | 1283 | 1704 | 2274 | 2744 | 2635 | 2819 | 3742 |
| 2004 | 559 | 1006 | 1258 | 1579 | 2044 | 2809 | 3123 | 2945 | 3759 |
| 2005 | 339 | 886 | 1265 | 1506 | 1916 | 2323 | 3028 | 3211 | 2891 |
| 2006 | 402 | 749 | 1093 | 1495 | 1758 | 2163 | 2555 | 3054 | 3589 |
| 2007 | 510 | 748 | 988 | 1346 | 1840 | 2062 | 2350 | 2525 | 3143 |
| 2008 | 383 | 636 | 857 | 1125 | 1575 | 2149 | 2417 | 2802 | 2600 |
| 2009 | 452 | 841 | 960 | 1131 | 1352 | 1757 | 2364 | 2497 | 3074 |
| 2010 | 447 | 756 | 1092 | 1294 | 1448 | 1685 | 2188 | 2366 | 2646 |
| 2011 | 588 | 905 | 1122 | 1455 | 1688 | 1914 | 2094 | 2455 | 2986 |
| 2012 | 668 | 978 | 1222 | 1492 | 1903 | 2164 | 2366 | 2704 | 2940 |
| 2013 | 678 | 1084 | 1358 | 1675 | 2036 | 2400 | 2554 | 3097 | 3097 |
| 2014 | 536 | 1080 | 1433 | 1793 | 2121 | 2504 | 2624 | 3178 | 3349 |
| 2015 | 573 | 1084 | 1486 | 2011 | 2332 | 2823 | 3306 | 3258 | 3768 |
| 2016 | 406 | 1049 | 1521 | 1940 | 2373 | 2529 | 3089 | 3370 | 3356 |
| 2017 | 437 | 826 | 1591 | 2056 | 2432 | 2795 | 2920 | 3352 | 3556 |

Table 10.1.6 Haddock in division Va Weight at age in the catches. Predicted values are shaded.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 620 | 960 | 1410 | 2030 | 2910 | 3800 | 4560 | 4720 | 5956 |
| 1980 | 837 | 831 | 1306 | 2207 | 2738 | 3188 | 3843 | 4506 | 4983 |
| 1981 | 584 | 693 | 1081 | 1656 | 2283 | 3214 | 3409 | 4046 | 5261 |
| 1982 | 289 | 959 | 1455 | 1674 | 2351 | 3031 | 3481 | 3874 | 4123 |
| 1983 | 320 | 1006 | 1496 | 1921 | 2371 | 2873 | 3678 | 4265 | 4502 |
| 1984 | 691 | 1007 | 1544 | 2120 | 2514 | 3027 | 2940 | 3906 | 4033 |
| 1985 | 652 | 1125 | 1811 | 2260 | 2924 | 3547 | 3733 | 4039 | 4659 |
| 1986 | 336 | 1227 | 1780 | 2431 | 2771 | 3689 | 3820 | 4258 | 4456 |
| 1987 | 452 | 1064 | 1692 | 2408 | 3000 | 3565 | 4215 | 4502 | 4025 |
| 1988 | 362 | 780 | 1474 | 2217 | 2931 | 3529 | 3781 | 4467 | 4418 |
| 1989 | 323 | 857 | 1185 | 1996 | 2893 | 4066 | 3866 | 4734 | 4990 |
| 1990 | 269 | 700 | 1054 | 1562 | 2364 | 3414 | 4134 | 4946 | 4451 |
| 1991 | 288 | 699 | 979 | 1412 | 1887 | 2674 | 3135 | 4341 | 4957 |
| 1992 | 313 | 806 | 1167 | 1524 | 1950 | 2357 | 3075 | 4053 | 4703 |
| 1993 | 303 | 705 | 1333 | 1875 | 2386 | 2996 | 3059 | 3363 | 4409 |
| 1994 | 337 | 668 | 1019 | 1717 | 2391 | 2717 | 3280 | 3156 | 3278 |
| 1995 | 351 | 746 | 1096 | 1318 | 2044 | 2893 | 3049 | 3675 | 3137 |
| 1996 | 311 | 787 | 1187 | 1560 | 1849 | 2670 | 3510 | 3567 | 3731 |
| 1997 | 379 | 764 | 1163 | 1649 | 1943 | 2342 | 3020 | 3337 | 3236 |
| 1998 | 445 | 724 | 1147 | 1683 | 2250 | 2475 | 2834 | 3333 | 3596 |
| 1999 | 555 | 908 | 1101 | 1658 | 2216 | 2659 | 2928 | 3209 | 3513 |
| 2000 | 495 | 978 | 1333 | 1481 | 2119 | 2696 | 3307 | 3597 | 3757 |
| 2001 | 541 | 945 | 1456 | 1731 | 1832 | 2243 | 3020 | 3328 | 4236 |
| 2002 | 564 | 928 | 1253 | 1737 | 2219 | 2230 | 2911 | 3365 | 4387 |
| 2003 | 498 | 922 | 1283 | 1704 | 2274 | 2744 | 2635 | 2819 | 3742 |
| 2004 | 559 | 1006 | 1258 | 1579 | 2044 | 2809 | 3123 | 2945 | 3759 |
| 2005 | 339 | 886 | 1265 | 1506 | 1916 | 2323 | 3028 | 3211 | 2891 |
| 2006 | 402 | 749 | 1093 | 1495 | 1758 | 2163 | 2555 | 3054 | 3589 |
| 2007 | 510 | 748 | 988 | 1346 | 1840 | 2062 | 2350 | 2525 | 3143 |
| 2008 | 383 | 636 | 857 | 1125 | 1575 | 2149 | 2417 | 2802 | 2600 |
| 2009 | 452 | 841 | 960 | 1131 | 1352 | 1757 | 2364 | 2497 | 3074 |
| 2010 | 447 | 756 | 1092 | 1294 | 1448 | 1685 | 2188 | 2366 | 2646 |
| 2011 | 588 | 905 | 1122 | 1455 | 1688 | 1914 | 2094 | 2455 | 2986 |
| 2012 | 668 | 978 | 1222 | 1492 | 1903 | 2164 | 2366 | 2704 | 2940 |
| 2013 | 678 | 1084 | 1358 | 1675 | 2036 | 2400 | 2554 | 3097 | 3097 |
| 2014 | 536 | 1080 | 1433 | 1793 | 2121 | 2504 | 2624 | 3178 | 3349 |
| 2015 | 573 | 1084 | 1486 | 2011 | 2332 | 2823 | 3306 | 3258 | 3768 |
| 2016 | 406 | 1049 | 1521 | 1940 | 2373 | 2529 | 3089 | 3370 | 3356 |
| 2017 | 437 | 826 | 1591 | 2056 | 2432 | 2795 | 2920 | 3352 | 3556 |

Table 10.1.7 Haddock in division Va Sexual maturity-at-age in the stock. (from the March survey). Predicted values are shaded. The numbers for age 10 only apply to the spawning stock.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1980 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1981 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1982 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1983 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1984 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1985 | 0.016 | 0.144 | 0.536 | 0.577 | 0.765 | 0.766 | 0.961 | 0.934 | 1 |
| 1986 | 0.021 | 0.205 | 0.413 | 0.673 | 0.845 | 0.884 | 0.952 | 0.986 | 1 |
| 1987 | 0.022 | 0.137 | 0.426 | 0.535 | 0.778 | 0.776 | 1 | 0.969 | 1 |
| 1988 | 0.013 | 0.221 | 0.394 | 0.767 | 0.793 | 0.928 | 0.914 | 1 | 1 |
| 1989 | 0.041 | 0.202 | 0.532 | 0.727 | 0.818 | 0.998 | 1 | 1 | 1 |
| 1990 | 0.114 | 0.334 | 0.634 | 0.814 | 0.843 | 0.918 | 0.882 | 1 | 1 |
| 1991 | 0.063 | 0.224 | 0.592 | 0.739 | 0.817 | 0.894 | 0.495 | 1 | 1 |
| 1992 | 0.05 | 0.227 | 0.419 | 0.799 | 0.901 | 0.901 | 0.858 | 1 | 1 |
| 1993 | 0.124 | 0.362 | 0.481 | 0.67 | 0.904 | 0.977 | 0.908 | 0.867 | 1 |
| 1994 | 0.248 | 0.312 | 0.573 | 0.762 | 0.846 | 1 | 0.907 | 1 | 1 |
| 1995 | 0.124 | 0.479 | 0.382 | 0.75 | 0.753 | 0.606 | 0.985 | 1 | 1 |
| 1996 | 0.191 | 0.362 | 0.59 | 0.648 | 0.787 | 0.739 | 0.949 | 0.908 | 1 |
| 1997 | 0.093 | 0.436 | 0.587 | 0.683 | 0.75 | 0.783 | 0.88 | 1 | 1 |
| 1998 | 0.026 | 0.454 | 0.668 | 0.77 | 0.733 | 0.849 | 0.899 | 1 | 1 |
| 1999 | 0.05 | 0.397 | 0.683 | 0.724 | 0.749 | 0.892 | 0.761 | 0.92 | 1 |
| 2000 | 0.107 | 0.261 | 0.632 | 0.808 | 0.868 | 0.873 | 1 | 0.78 | 1 |
| 2001 | 0.091 | 0.377 | 0.522 | 0.753 | 0.895 | 0.916 | 0.918 | 1 | 1 |
| 2002 | 0.047 | 0.286 | 0.633 | 0.8 | 0.934 | 0.928 | 1 | 1 | 1 |
| 2003 | 0.062 | 0.347 | 0.685 | 0.867 | 0.922 | 0.946 | 1 | 1 | 1 |
| 2004 | 0.037 | 0.361 | 0.57 | 0.831 | 0.91 | 1 | 1 | 1 | 1 |
| 2005 | 0.024 | 0.23 | 0.562 | 0.753 | 0.927 | 0.936 | 0.968 | 1 | 1 |
| 2006 | 0.027 | 0.117 | 0.462 | 0.621 | 0.739 | 0.918 | 1 | 1 | 1 |
| 2007 | 0.078 | 0.208 | 0.418 | 0.68 | 0.77 | 0.875 | 0.959 | 1 | 1 |
| 2008 | 0.027 | 0.263 | 0.418 | 0.621 | 0.828 | 0.87 | 0.904 | 0.975 | 1 |
| 2009 | 0.017 | 0.301 | 0.47 | 0.576 | 0.847 | 0.891 | 1 | 0.968 | 1 |
| 2010 | 0.029 | 0.187 | 0.618 | 0.778 | 0.787 | 0.887 | 0.934 | 1 | 0.958 |
| 2011 | 0.045 | 0.176 | 0.426 | 0.823 | 0.816 | 0.838 | 0.899 | 0.974 | 1 |
| 2012 | 0.106 | 0.167 | 0.445 | 0.627 | 0.819 | 0.903 | 0.852 | 0.911 | 1 |
| 2013 | 0.046 | 0.223 | 0.381 | 0.714 | 0.793 | 0.92 | 0.986 | 0.974 | 0.992 |
| 2014 | 0.107 | 0.192 | 0.391 | 0.567 | 0.675 | 0.735 | 0.925 | 0.906 | 0.883 |
| 2015 | 0.138 | 0.283 | 0.445 | 0.667 | 0.795 | 0.772 | 0.892 | 1 | 0.889 |
| 2016 | 0.067 | 0.417 | 0.696 | 0.827 | 0.894 | 0.912 | 0.935 | 1 | 1 |
| 2017 | 0.066 | 0.317 | 0.694 | 0.835 | 0.894 | 0.925 | 0.935 | 1 | 1 |

Table 10.2.1 Haddock in division Va. Summary table from the SPALY run using the surveys in March and October for tuning.

| Year | ReCRUITMENT THOUSAND AT AGE 2 | BIOMASS $3+$ TONS | SSB TONS | LANDINGS TONS | Yield/SSB | F4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 80923 | 162177 | 96072 | 55330 | 0.521 | 0.576 |
| 1980 | 37390 | 192244 | 116521 | 51110 | 0.398 | 0.439 |
| 1981 | 10426 | 206988 | 141628 | 63558 | 0.542 | 0.449 |
| 1982 | 42788 | 180380 | 136817 | 69428 | 0.444 | 0.507 |
| 1983 | 29306 | 148112 | 112589 | 65942 | 0.508 | 0.586 |
| 1984 | 20574 | 112797 | 82961 | 48282 | 0.515 | 0.582 |
| 1985 | 42788 | 102394 | 66652 | 51102 | 0.537 | 0.767 |
| 1986 | 86501 | 96480 | 59837 | 48859 | 0.739 | 0.817 |
| 1987 | 164036 | 105395 | 46298 | 40760 | 0.584 | 0.88 |
| 1988 | 48742 | 153708 | 69391 | 54204 | 0.675 | 0.781 |
| 1989 | 29778 | 168184 | 99537 | 62885 | 0.676 | 0.632 |
| 1990 | 27094 | 145507 | 110745 | 67198 | 0.611 | 0.607 |
| 1991 | 92280 | 122708 | 89825 | 54692 | 0.664 | 0.609 |
| 1992 | 175094 | 106310 | 66379 | 47121 | 0.728 | 0.71 |
| 1993 | 38437 | 130461 | 71000 | 48123 | 0.669 | 0.678 |
| 1994 | 46842 | 127836 | 83295 | 59502 | 0.641 | 0.714 |
| 1995 | 72857 | 124042 | 85054 | 60884 | 0.661 | 0.716 |
| 1996 | 36341 | 108036 | 70008 | 56890 | 0.675 | 0.813 |
| 1997 | 102509 | 87152 | 58993 | 43764 | 0.624 | 0.742 |
| 1998 | 17976 | 97121 | 64203 | 41192 | 0.627 | 0.642 |
| 1999 | 50160 | 91024 | 64439 | 45411 | 0.685 | 0.705 |
| 2000 | 117423 | 90674 | 63509 | 42105 | 0.636 | 0.663 |
| 2001 | 156535 | 115046 | 70366 | 39654 | 0.462 | 0.564 |
| 2002 | 187267 | 168427 | 99344 | 50498 | 0.461 | 0.508 |
| 2003 | 50394 | 219757 | 147523 | 60883 | 0.404 | 0.413 |
| 2004 | 151137 | 252826 | 181303 | 84828 | 0.491 | 0.468 |
| 2005 | 384765 | 258912 | 176994 | 97225 | 0.522 | 0.549 |
| 2006 | 90617 | 298798 | 143410 | 97614 | 0.577 | 0.681 |
| 2007 | 42783 | 297360 | 162516 | 109966 | 0.555 | 0.677 |
| 2008 | 44466 | 249662 | 158368 | 102872 | 0.475 | 0.65 |
| 2009 | 121069 | 192792 | 142494 | 82045 | 0.504 | 0.576 |
| 2010 | 41838 | 168658 | 114084 | 64168 | 0.469 | 0.562 |
| 2011 | 32441 | 153885 | 97987 | 49433 | 0.402 | 0.504 |
| 2012 | 21413 | 144883 | 95026 | 46208 | 0.331 | 0.486 |
| 2013 | 40472 | 138496 | 100228 | 44097 | 0.324 | 0.44 |
| 2014 | 27875 | 125286 | 77383 | 33900 | 0.275 | 0.438 |
| 2015 | 14093 | 127585 | 87450 | 39646 | 0.373 | 0.453 |
| 2016 | 119694 | 109241 | 76722 |  |  |  |
| $\begin{gathered} \text { Average } \\ \text { 1979-2015 } \end{gathered}$ | 76240 | 154772 | 99657 | 58287 | 0.535 | 0.606 |

Table 10.2.2 Haddock in division Va. Number in stock from the SPALY run using both the surveys. Shaded cells are input to prediction. . Predictions shown are based on HCR.

| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 46 | 81 | 117.3 | 27.7 | 19.6 | 20.44 | 3.41 | 0.77 | 0.15 | 0.05 |
| 1980 | 13 | 37 | 66.1 | 94.3 | 19.3 | 10.54 | 8.57 | 1.21 | 0.23 | 0.07 |
| 1981 | 52 | 10 | 30.1 | 52.9 | 66.8 | 11.91 | 5.19 | 3.64 | 0.5 | 0.11 |
| 1982 | 36 | 43 | 8.5 | 24.2 | 38.9 | 39.42 | 4.33 | 1.69 | 1.35 | 0.26 |
| 1983 | 25 | 29 | 34.9 | 6.8 | 16.9 | 21.99 | 19.56 | 1.67 | 0.48 | 0.45 |
| 1984 | 52 | 21 | 24 | 27.7 | 4.1 | 9.7 | 9.09 | 8.02 | 0.68 | 0.21 |
| 1985 | 106 | 43 | 16.6 | 18.6 | 18.2 | 2.14 | 3.63 | 4.06 | 2.88 | 0.34 |
| 1986 | 200 | 86 | 33.8 | 12.1 | 11.1 | 8.75 | 0.98 | 1.45 | 1.59 | 0.63 |
| 1987 | 60 | 164 | 69.9 | 23.9 | 6.2 | 4.88 | 2.51 | 0.35 | 0.46 | 0.49 |
| 1988 | 36 | 49 | 132.6 | 49.7 | 13.2 | 2.59 | 2.15 | 1.07 | 0.17 | 0.23 |
| 1989 | 33 | 30 | 39.7 | 99.6 | 27 | 5.58 | 0.94 | 0.78 | 0.32 | 0.08 |
| 1990 | 113 | 27 | 24.2 | 30.3 | 61.1 | 13.43 | 1.68 | 0.31 | 0.14 | 0.13 |
| 1991 | 214 | 92 | 20.5 | 17.6 | 17 | 28.7 | 5.27 | 0.63 | 0.12 | 0.05 |
| 1992 | 47 | 175 | 67.8 | 14.8 | 9.6 | 7.25 | 10.74 | 1.92 | 0.21 | 0.06 |
| 1993 | 57 | 38 | 138.5 | 45.8 | 7 | 3.65 | 2.59 | 4.08 | 0.67 | 0.07 |
| 1994 | 89 | 47 | 30.8 | 102.2 | 25.9 | 3.03 | 1.43 | 0.83 | 1.31 | 0.23 |
| 1995 | 44 | 73 | 35.5 | 22.2 | 58.1 | 11.54 | 1.15 | 0.52 | 0.36 | 0.49 |
| 1996 | 125 | 36 | 56 | 22.4 | 12.9 | 25.93 | 4.23 | 0.38 | 0.16 | 0.13 |
| 1997 | 22 | 102 | 27 | 36.3 | 11.4 | 6.56 | 9.56 | 1.34 | 0.13 | 0.06 |
| 1998 | 61 | 18 | 82.7 | 18.5 | 19.9 | 4.93 | 3.01 | 3.27 | 0.45 | 0.05 |
| 1999 | 143 | 50 | 14.5 | 60.2 | 10 | 9.15 | 1.85 | 1.18 | 0.96 | 0.13 |
| 2000 | 191 | 117 | 39.8 | 10.4 | 33.7 | 4.12 | 3.11 | 0.69 | 0.4 | 0.35 |
| 2001 | 229 | 156 | 93.7 | 26.9 | 6.4 | 15.28 | 1.61 | 1.01 | 0.27 | 0.13 |
| 2002 | 62 | 187 | 125.9 | 66.7 | 15.7 | 3.95 | 6.85 | 0.71 | 0.39 | 0.12 |
| 2003 | 185 | 50 | 152.3 | 93.6 | 40.1 | 8.23 | 2.21 | 2.76 | 0.34 | 0.17 |
| 2004 | 470 | 151 | 40.9 | 119 | 61.9 | 21.5 | 4.05 | 1.13 | 1.14 | 0.2 |
| 2005 | 111 | 385 | 122.3 | 29.8 | 81.4 | 33.19 | 9.58 | 1.56 | 0.5 | 0.49 |
| 2006 | 52 | 91 | 312.8 | 91.6 | 18.1 | 43.68 | 14.71 | 3.7 | 0.55 | 0.18 |
| 2007 | 54 | 43 | 74 | 247.1 | 55.8 | 8.83 | 18.71 | 5.17 | 1.05 | 0.16 |
| 2008 | 148 | 44 | 34.3 | 57 | 164.1 | 25.52 | 4.2 | 6.31 | 1.75 | 0.39 |
| 2009 | 51 | 121 | 34.3 | 24 | 37.9 | 86.35 | 10.93 | 1.88 | 2.43 | 0.7 |
| 2010 | 40 | 42 | 98.2 | 25.2 | 15.2 | 23.01 | 39.02 | 3.76 | 0.88 | 0.74 |
| 2011 | 26 | 32 | 34.1 | 74.9 | 14.2 | 8.13 | 12.71 | 16.11 | 1.38 | 0.4 |
| 2012 | 49 | 21 | 26.3 | 26.5 | 50.6 | 7.07 | 4.08 | 6.8 | 7.56 | 0.69 |
| 2013 | 34 | 40 | 17.4 | 20.4 | 18.6 | 29.61 | 3.78 | 2.22 | 3.33 | 3.78 |
| 2014 | 17 | 28 | 32.9 | 13.3 | 14.1 | 11.62 | 15.9 | 2 | 1.22 | 1.59 |
| 2015 | 146 | 14 | 22.6 | 25.6 | 9.3 | 9.06 | 7.03 | 8.73 | 1.23 | 0.65 |
| 2016 | 66 | 120 | 11.3 | 17.1 | 17.2 | 5.5 | 4.78 | 3.39 | 4.69 | 0.81 |
| 2017 | 67 | 54 | 98 | 8.2 | 11 | 10.08 | 3.01 | 2.56 | 1.78 | 2.46 |
| 2018 | 67 | 55 | 43.8 | 74.1 | 5 | 5.93 | 5.09 | 1.46 | 1.24 | 0.86 |

Table 10.2.3 Haddock in division Va. Fishing mortality from the SPALY run using the March and October surveys for tuning. Predictions based on F4-7 $=0.3$ are highlighted.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 0.002 | 0.018 | 0.162 | 0.419 | 0.669 | 0.833 | 0.99 | 0.553 | 0 |
| 1980 | 0.018 | 0.023 | 0.144 | 0.282 | 0.508 | 0.657 | 0.685 | 0.561 | 0.724 |
| 1981 | 0.001 | 0.019 | 0.108 | 0.328 | 0.813 | 0.92 | 0.793 | 0.463 | 0.569 |
| 1982 | 0.003 | 0.032 | 0.156 | 0.369 | 0.501 | 0.751 | 1.056 | 0.903 | 1.288 |
| 1983 | 0.001 | 0.032 | 0.301 | 0.357 | 0.683 | 0.692 | 0.706 | 0.643 | 1.051 |
| 1984 | 0.013 | 0.051 | 0.22 | 0.449 | 0.784 | 0.607 | 0.825 | 0.493 | 0.369 |
| 1985 | 0.035 | 0.122 | 0.315 | 0.532 | 0.582 | 0.719 | 0.737 | 1.314 | 1.184 |
| 1986 | 0.013 | 0.148 | 0.467 | 0.625 | 1.048 | 0.816 | 0.937 | 0.976 | 0.918 |
| 1987 | 0.013 | 0.141 | 0.389 | 0.669 | 0.62 | 0.657 | 0.53 | 0.5 | 0.685 |
| 1988 | 0.005 | 0.086 | 0.411 | 0.665 | 0.811 | 0.815 | 0.998 | 0.557 | 0.557 |
| 1989 | 0.007 | 0.071 | 0.288 | 0.498 | 1.003 | 0.917 | 1.552 | 0.682 | 0.632 |
| 1990 | 0.079 | 0.117 | 0.379 | 0.556 | 0.736 | 0.772 | 0.769 | 0.794 | 0.467 |
| 1991 | 0.109 | 0.123 | 0.413 | 0.651 | 0.783 | 0.811 | 0.89 | 0.473 | 0.25 |
| 1992 | 0.035 | 0.192 | 0.555 | 0.762 | 0.827 | 0.768 | 0.858 | 0.973 | 0.204 |
| 1993 | 0.022 | 0.104 | 0.37 | 0.635 | 0.736 | 0.934 | 0.933 | 0.842 | 0.383 |
| 1994 | 0.078 | 0.128 | 0.365 | 0.608 | 0.769 | 0.821 | 0.643 | 0.786 | 0.575 |
| 1995 | 0.063 | 0.259 | 0.337 | 0.607 | 0.804 | 0.895 | 0.971 | 0.856 | 0.926 |
| 1996 | 0.099 | 0.233 | 0.473 | 0.48 | 0.798 | 0.95 | 0.912 | 0.79 | 0.756 |
| 1997 | 0.015 | 0.176 | 0.404 | 0.641 | 0.579 | 0.873 | 0.9 | 0.819 | 0.253 |
| 1998 | 0.017 | 0.117 | 0.413 | 0.575 | 0.781 | 0.738 | 1.025 | 1.041 | 0.53 |
| 1999 | 0.032 | 0.126 | 0.38 | 0.689 | 0.878 | 0.792 | 0.87 | 0.806 | 0.776 |
| 2000 | 0.025 | 0.193 | 0.286 | 0.591 | 0.737 | 0.93 | 0.74 | 0.933 | 0.807 |
| 2001 | 0.018 | 0.14 | 0.337 | 0.286 | 0.603 | 0.62 | 0.745 | 0.568 | 0.44 |
| 2002 | 0.006 | 0.096 | 0.308 | 0.445 | 0.381 | 0.71 | 0.523 | 0.65 | 0.468 |
| 2003 | 0.009 | 0.047 | 0.213 | 0.424 | 0.508 | 0.469 | 0.685 | 0.345 | 0.383 |
| 2004 | 0.012 | 0.116 | 0.179 | 0.424 | 0.609 | 0.753 | 0.616 | 0.645 | 0.71 |
| 2005 | 0.007 | 0.089 | 0.297 | 0.423 | 0.614 | 0.753 | 0.849 | 0.809 | 0.653 |
| 2006 | 0.003 | 0.036 | 0.296 | 0.519 | 0.648 | 0.846 | 1.056 | 1.057 | 0.829 |
| 2007 | 0.02 | 0.06 | 0.209 | 0.581 | 0.544 | 0.886 | 0.882 | 0.787 | 0.58 |
| 2008 | 0.059 | 0.157 | 0.208 | 0.442 | 0.648 | 0.602 | 0.757 | 0.723 | 0.636 |
| 2009 | 0.01 | 0.108 | 0.255 | 0.3 | 0.594 | 0.867 | 0.555 | 0.992 | 0.987 |
| 2010 | 0.003 | 0.07 | 0.37 | 0.428 | 0.393 | 0.685 | 0.8 | 0.584 | 0.963 |
| 2011 | 0.009 | 0.053 | 0.191 | 0.501 | 0.491 | 0.425 | 0.557 | 0.501 | 0.366 |
| 2012 | 0.01 | 0.057 | 0.154 | 0.337 | 0.427 | 0.406 | 0.515 | 0.493 | 0.491 |
| 2013 | 0.007 | 0.069 | 0.169 | 0.272 | 0.422 | 0.435 | 0.403 | 0.535 | 0.51 |
| 2014 | 0.009 | 0.051 | 0.158 | 0.241 | 0.302 | 0.4 | 0.283 | 0.433 | 0.521 |
| 2015 | 0.018 | 0.078 | 0.198 | 0.323 | 0.44 | 0.531 | 0.421 | 0.226 | 0.672 |
| 2016 | 0 | 0.117 | 0.241 | 0.335 | 0.402 | 0.423 | 0.444 | 0.444 | 0.444 |
| 2017 | 0.001 | 0.08 | 0.306 | 0.419 | 0.483 | 0.524 | 0.524 | 0.524 | 0.524 |
| 2018 | 0.003 | 0.098 | 0.25 | 0.467 | 0.542 | 0.568 | 0.568 | 0.568 | 0.568 |

Table 10.4.1 Output from short-term predictions. Numbers here apply to calendar years. The adopted HCR lead to TAC of 34.6 kt for the fishing year 2016/2017.

| 2016 |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Bio 3+ | SSB | Fmult | F4-7 | Landings |
| 109 | 77 | 0.955 | 0.356 | 34 |


|  |  | 2017 |  |  |  | 2018 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FмULT | F4-7 | BıO 3+ | SSB | LANDINGS | Bıo 3+ | SSB |
| 0.1 | 0.037 | 129 | 88 | 4 | 177 | 127 |
| 0.2 | 0.075 | 129 | 88 | 7 | 173 | 124 |
| 0.3 | 0.112 | 129 | 88 | 11 | 170 | 121 |
| 0.4 | 0.149 | 129 | 88 | 14 | 167 | 118 |
| 0.5 | 0.186 | 129 | 88 | 18 | 163 | 115 |
| 0.6 | 0.224 | 129 | 88 | 21 | 160 | 112 |
| 0.7 | 0.261 | 129 | 88 | 24 | 157 | 110 |
| 0.8 | 0.298 | 129 | 88 | 27 | 154 | 107 |
| 0.9 | 0.335 | 129 | 88 | 30 | 151 | 105 |
| 1 | 0.373 | 129 | 88 | 33 | 148 | 102 |
| 1.1 | 0.41 | 129 | 88 | 35 | 146 | 100 |
| 1.2 | 0.447 | 129 | 88 | 38 | 143 | 98 |
| 1.3 | 0.484 | 129 | 88 | 41 | 141 | 96 |
| 1.4 | 0.522 | 129 | 88 | 43 | 138 | 94 |
| 1.5 | 0.559 | 129 | 88 | 45 | 136 | 92 |
| 1.6 | 0.596 | 129 | 88 | 48 | 134 | 90 |
| 1.7 | 0.633 | 129 | 88 | 50 | 131 | 88 |
| 1.8 | 0.671 | 129 | 88 | 52 | 129 | 86 |
| 1.9 | 0.708 | 129 | 88 | 54 | 127 | 84 |
| 2 | 0.745 | 129 | 88 | 56 | 125 | 83 |



Figure 10.1.1 Haddock in division Va. Landings 1905-2015


Figure 10.1.2 Haddock Division VA. Landings in tons and percent of total by gear and year.


Figure 10.1.3 Haddock Division VA. Spatial distribution of landings. The legend show tonnes per square mile.


Figure 10.1.4 Haddock in division Va. Age disaggregated catch in tons.


Figure 10.1.5 Haddock in division Va. Percent of catch in tonnes 2015 (red) compared to last year's predictions.


Figure 10.1.6 Icelandic haddock. Total biomass indices from the groundfish surveys in March (lines and shading) and the groundfish survey in October vertical segments. The standard error in the estimate of the indices is shown in the figure. Due to a strike the autumn survey was not conducted in October 2011.

*Figure 10.1.7. Spatial distribution of haddock in the groundfish survey in March. The legend show kg per hour towed.


Figure 10.1.8. Proportion of the landings and the biomass of 42 cm and larger haddock that is in the north area. The small figure shows the northern area.


Figure 10.1.9 Haddock in division Va. Mean weight at age in the survey. Predictions are shown as red. The values shown are used as weight at age in the stock and spawning stock.


Figure 10.1.10 Haddock in division Va. Mean weight at age in the catches. Predictions are shown as red.


Figure 10.1.11 Haddock in division Va. Maturity-at-age in the survey. The blue bar indicates predictions. The values are used to calculate the spawning stock


Figure 10.1.12. Catch per unit of effort in the most important gear types. The bars are based on locations where more than $50 \%$ of the catch is haddock and the lines on all records where haddock is caught. A change occurred in the longline fleet starting September 1999. Earlier only vessels larger than 10 BRT were required to return logbooks but later all vessels were required to return logbooks. Not updated


Figure 10.2.1 Haddock in division Va. Summary from assessment. Red colours in lower figure indicates predicted values.


Figure 10.2.2. Haddock in division Va. Residuals from the fit to survey data from Adapt run based on the both the surveys. Coloured circles indicate positive residuals (observed $>$ modelled). The largest circle corresponds to a value of 0.87 . Residuals are proportional to the area of the circles. Lage efri harvest ratio


Figure 10.2.3. Haddock in division Va. Observed and predicted biomass from the surveys according to the SPALY run.


Figure 10.2.4. Haddock in division Va. Results from the SPALY run. Catchability and CV from the autumn survey (wide lines) and March survey (thinner lines). Estimates from 2008 shown dashed.



Figure 10.2.5. Haddock in division Va .Upper picture shows retrospective pattern of biomass 3+ from the SPALY run. Each retro ends 2 years after the assessment year i.e after the advisory year. The lower picture shows scaled retrospective pattern of fishing mortality. Average of Mohns rho over the last 7 years is $\mathbf{- 0 . 0 0 5}$. Errors in prediction of weight and maturity-at-age are not included and the assessment converges slowly.

10.2.6 Haddock in division Va. Estimate of the reference biomass 45 cm and larger from some different assessment models and tuning data. (SMB refers to March survey, SMH autumn survey and SMX both.

10.2.. 7 Haddock in division Va. Comparison of some of the results of 2016 assessment based on different tuning data and 2015 assessment tuned with both the surveys. .


Figure 10.2.8. Comparison of 2015 and 2016 assessment


Figure 10.3.1. Average long term spawning stock as function of harvest rate. Results are based on simulations without trigger point, autocorrelation of recruitment and assessment error.


Figure 10.3.1. Distribution of harvest rate in the long term based on the Harvest Control Rule simulations done in 2013.

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Byggt á gögnum Fiskistofu um landanir



Figure 10.4.1 Haddock in division Va. Development of the landings during the fishing year 2015/16 (left side) and calendar year (2016) on the right. Fishing year 2014/2015 and calendar year 2015 shown for comparison. Tac (kvóti) for the fishing year shown in the left figure.


Figure 10.4.2 Haddock in division Va. Input data to prediction. Predictions are based on the period since 2000. . Exponential of the yearfactor (growth multiplier) in the equation

$$
\log \frac{W_{a+1, t+1}}{W_{a, t}}=\alpha+\beta \log W_{a, t}+\delta_{y e a r}
$$



Figure 10.4.3 Haddock in division Va. Proportion of the biomass of a yearclass above certain size. The points show data, compiled from the March survey and the lines a curve fitted to the data and used in simulations.

## 11 Icelandic summer spawning herring

### 11.1 Scientific data

### 11.1.1 Surveys description

The scientific data used for assessment of the Icelandic summer-spawning herring stock are based on annual acoustic surveys (IS-Her-Aco-4Q/1Q), which have been ongoing since 1974 (Table 11.1.1.1). These surveys have been conducted in October-January, and even as late as March. The surveyed area each year is decided on basis of available information on the distribution of the stock in previous and the current year, which include information from the fishery. Thus, the survey area varies spatially as the survey is focused on the adult and incoming year classes but is considered to cover the whole stock each year.

The acoustic abundance index for the adult stock in the winter 2015/2016 derives from two dedicated acoustic surveys in January and March 2016 (Óskarsson 2016). The nursery grounds of the stock were then covered on RV Dröfn in three surveys in September and October 2015. In addition to getting an acoustic estimate on the adult part and on juveniles at age 1, the objective was also to get an estimate of prevalence of Ichthyophonus infection in the stock. The instrument and methods in the surveys were the same as in previous years and described in the stock annex and all the results are detailed in a WD to NWWG (Óskarsson 2016). The biological sampling in the survey is detailed in Table 11.1.1.2.

### 11.1.2 The surveys' results

The fishable part of the herring stock was observed in three main areas, in an offshore area west of Iceland (Kolluáll and Jökuldjúp), south of Iceland around Vestmannaeyjar, and in offshore areas east of Iceland (Litladýpi og Reyðarfjarðardjúp) (Figure 11.1.2.1; Óskarsson 2016). As in previous years, most of the stock was measured acoustically west of Iceland (in Kolluáll) and to a smaller and decreasing degree east of Iceland (Figure 11.1.2.2.). The total amount of the adult stock (age $4^{+}$) came to 372 kt , compare to 433 kt in the autumn 2014. The total biomass of the fishable stock $(>26 \mathrm{~cm})$ was 396 kt compared with 450 kt in the autumn 2014, which represents $12 \%$ decrease in biomass while there was a $7 \%$ decrease in number ( $1411 \times 10^{6}$ compare to $1514 \times 10^{6}$, respectively).

The 2010 year class (age 5 in the autumn 2015) was the most numerous in the survey or $23 \%$ of the total number of herring (Table 11.1.1.1). The 2008 and 2009 year classes were also numerous ( 16 and $15 \%$, respectively), and together these year classes contributed to $53 \%$ of the total number and biomass.

The number of juvenile herring (i.e. age 1) observed acoustically amounted to 382 million fish. Applying the linear-regression provided by Gudmundsdottir et al. (2007) implied that the 2014 year class will be 391 million at age 3 in the autumn 2017, or below average year class size ( 666 million at age 3 and geometric mean of 588 millions). This number is used in the forecast in the 2016 assessment below.

As discussed in Óskarsson (2016), the acoustic measurements in January east of Iceland are likely to have higher uncertainty than elsewhere and normally in this survey, because of sparse distribution of relatively large schools (i.e. only three schools provided the acoustic values), large distance between survey tracks, limited spatial coverage of the survey in the area and bad weather condition.

### 11.1.3 Prevalence of /chthyophonus infection in the stock

In a working document to NWWG 2013, Óskarsson and Pálsson (2013) addressed the development and nature of the massive and long-lasting Ichthyophonus hoferi outbreak in Icelandic summer-spawning herring since the autumn 2008-2013. Their main conclusions were that the infection was only causing significant additional mortality in the first two years, despite a high prevalence of infection for five years. It indicated that the infection to be less lethal for herring than had been assumed in previous assessments. This was followed in the 2013, 2014 and 2015 assessments (ICES 2013; 2014; 2015), where additional natural mortality because of the infection, and estimated from catch samples (e.g. Óskarsson et al. 2012; ICES 2012), was only be applied for the years 2009 and 2010, but not the following years.

The results of this year's investigations are supporting this main conclusion of not significant infection mortality since 2010.

The prevalence of infection in the Icelandic summer-spawning herring in the winter 2015/2016 west of Iceland was highest for the 2005 and 2003 year classes, or $43 \%$ and $37 \%$, respectively (Figure 11.1.3.1). The 2004, 2006 and 2007 year classes were not far off with prevalence of infection $30-33 \%$ this winter. Thus this year's results are in agreement with previous results as the prevalence is still high for these year classes and at comparable levels to previous years. However, the prevalence of infection has been increasing for the recent two years in the 2007 and 2008 year classes, and for the last year in the 2009 year class. Moreover, two years old herring at the nursery grounds north of Iceland in the autumn 2015 had low level of prevalence of infection, which has not has been observed there since the autumn 2011 (Óskarsson and Pálsson 2016). This together indicates that some new infection, even if at much lower rate than in 20082009, has been taking place in the most recent two years.
Further work is still ongoing, including analysing the difference in fit (i.e. comparing the residuals sum of square) of the assessment model used for the stock (NFT-Adapt) by varying the M related to infection. The stable high level of prevalence of infection in the older age groups (year classes 2002-2006) for such a long period and they still being fishable (Figure 11.1.3.1), supports that the mortality rate caused by the infection is very low and insignificant. Thus, increase in prevalence of infection for year classes 2007-2009 does not necessarily, and probably not, mean induced mortality by infection in the near future. However, it calls for continuation of monitoring the situation closely.

### 11.2 Information from the fishing industry

The total landings of Icelandic summer-spawning herring in 2015/2016 season were about 69.7 kt with no discards reported (Table 11.2.1 and in Figure 11.2.1). Note that the total landings include also bycatches in the mackerel fishery in June-August 2015 $(4.5 \mathrm{kt})$, even if they belong to the official fishing season 2014/2015. This is a traditional method in assessment of the stock. The quality of the herring landing data regarding discards and misreporting is consider to be adequate as implied in a general summary in section 7 and in the Her-Vasu stock annex. The recommended TAC, provided in the spring 2015, was 71 kt and allowable TAC 68 kt because of 3 kt overshoot in the season before that was transferred to the next season.

The direct fishery started in October in offshore areas west of Iceland. Most of the catches were taken over a wide area there in October to January in pelagic trawls, or $89 \%$ of the total catch (Figure 11.2.2). In October and November 2.4 kt were caught east of Iceland (3\%), while the remaining catch of $5.5 \mathrm{kt}(8 \%)$ was taken as bycatch in the
fishery for the Norwegian spring-spawning herring, NSSH, and Atlantic mackerel during June to September.
Like in some of the previous winters, spring-spawning herring (Icelandic spring spawners or NSSH) was mixed with the Icelandic summer-spawning herring stock in the catches in the winter 2015/2016. This applied to the fishery in the east in October and November where $24 \%$ was NSSH according to four catch samples, and consequently reported as NSSH. Based on maturity stage of the herring in catch samples, $2.4 \%$ of the herring caught west of Iceland in the winter 2015/2016 were spring spawners.

### 11.2.1 Fleets and fishing grounds

The herring fishing season has taken minor changes in the last three decades as detailed in the stock annex. All seasonal restricted landings, catches and recommended TACs since 1984 are given in thousands tonnes (kt) in Table 11.2.1.
Around $97 \%$ of the catch in 2015/2016 was taken in pelagic trawls and $3 \%$ in purseseine (Figure 11.2.1), which reflects that both the targeting and bycatch fisheries take mainly place in offshore areas. No driftnets fishery took place in Breiðafjörður as in the most recent four years. During all fishing seasons since 2007/2008 to 2012/2013, most of the catches ( $\sim 90 \%$ ) were been taken west off Iceland in Breiðafjörður, while prior to that they were mainly taken off the south-, southeast-, and the east coast. In 2013/2014 there was an indication for changes in this pattern, with less proportion in Breiðafjörður, and then in 2014/2015 almost all of the overwintering west of Iceland took place offshore which continued this winter. These changes in distribution explain the dominance of pelagic trawl in the fishery, which is preferred by the fleet over purse-seine in offshore areas.

To protect juveniles herring ( 27 cm and smaller) in the fishery, area closures are enforced based on a regulation of the herring fishery set by the Icelandic Ministry of Fisheries (no. 376, 8. Oktober 1992). No closure was enforced in this herring fishery in $2015 / 16$. Normally, the age of first recruitment to the fishery is age- 3 , which is fish at length around $26-29 \mathrm{~cm}$.

### 11.2.2 Catch in numbers, weight at age and maturity

## Catch-at-age in 2015/2016:

The procedure for the catch-at-age estimations, as described in the Stock Annex, was followed for the 2015/16 fishing season. It involves calculations from catch data collected at the harbours by the research personnel $(0 \%)$ or at sea by fishers $(100 \%)$. This year, the calculations were accomplished by dividing the total catch into five cells confined by season and area as detailed in Óskarsson and Pálsson (2016). In the same way, five weight-at-length relationships derived from the length and weight measurements of the catch samples were used. On basis of difference in length-at-age between the summer months (June-Sept.) and the winter (Oct.-Jan.), two length-age keys were applied. The catches of the Icelandic summer spawners in number-at-age for this fishing season as well as back to 1982 are given in Table 11.2.2.1. The geographical location of the sampling is shown on Figure 11.2.2.

The age composition of the total catches in 2015/2016 was somewhat different from the composition in the bycatch of herring in the mackerel and NSS-herring fishery in the summer 2015 (Figure 11.2.2.1). The summer fishery included to a higher degree younger age groups (age $3-5 ; 48 \%$ of the biomass) than the whole fishery ( $38 \%$ ), and
consequently vice versa for older age groups. This difference is probably reflecting the geographical distribution of the different age groups, with larger proportion of these age groups in the east and south than in the west, according to the acoustic surveys ( $57 \% \mathrm{vs} .32 \%$ by biomass; Óskarsson 2016), where the main bycatch takes place.

## Weight at age:

As stated in the stock annex, the mean weight-at-age of the stock is derived from the catch samples (Table 11.2.2.2). The total number of fish weighed from the catch in 2015/16 was 2537 and 1674 of them were aged from their fish scales.

## Proportion mature:

The fixed maturity ogives were used in this year's assessment, as introduced in the stock annex, where proportion mature-at-age 3 is set $20 \%$ and $85 \%$ for fish at age 4 , while all older fish is considered mature.
Observed vs. predictions of catch composition:
The relative contribution of the different year (age) classes was similar to what was predicted in the analytical assessment in 2015 for age $5+$ while age 3 was less and age 4 more seen in the catches than predicted (Figure 11.2.2.1). Again, this reflects both the difference in geographical location of the fishery and the stock (age 3 mainly in the east and south; Óskarsson 2016) but also that the 2011 year class (age 4) is maybe not as extremely small as predicted in the 2015 assessment (ICES 2015).

### 11.3 Analytical assessment

### 11.3.1 Analysis of input data

Examination of catch curves for the year classes from 1984 to 2011 (Figure 11.3.1.1) indicates, in general, that the total mortality signal $(Z)$ in the fully recruited age groups is around 0.4 . It is under the assumption that the effort has been the same the whole time. In recent years the effort has changed a lot because of the infection and spatial distribution of the stock, and the mass mortality in 2012/2013, which makes any strong deductions from the catch curves for those recent less meaningful.

Catch curves were also plotted using the age disaggregated survey indices for each year class from 1984-2011 (Figure 11.3.1.2). Even if the total mortalities look at bit noisy for some year classes, they seem to be fairly close to 0.4 , for example for 1998-2005 year classes. There is an indication that the fish is fully assessable to the survey at age 3, but apparently a year later occasionally.

Mortality in the stock because of the Ichthyophonus outbreak cannot be detected clearly from the catch curves of the surveys. There is possibly a small change in level of the curve around 2009 for the big 1999 year classes. However, it should be noted that the highest prevalence of infection has been in the 2004, 2005 and 2006 year classes and they were not all fully in the survey prior to the infection outbreak. Further work on this matter is ongoing.

### 11.3.2 Exploration of different assessment models

In order to explore the data this year, two assessment tools were used, NFT-ADAPT (VPA/ADPAT version 3.3.0 NOAA Fisheries Toolbox) and TASACS (VPA module; Skagen and Skålevik 2009). The NFT-Adapt has been used as the basis for the assessments since 2005 and it was considered appropriate as the principal assessment tool for the stock at benchmark assessment in January 2011 (ICES 2011a). The VPA module
in TASACS was, on the other hand, the principal assessment model for Norwegian spring-spawning herring during 2008-2015 (ICES 2016) and a familiar tool to the assessment scientists. The catch data used were from 1987/88-2015/16 (Table 11.2.2.1) and survey data from 1987/88-2015/16 (Table 11.1.1.1) for NFT-Adapt but back to 1975 for TASACS (see below). Other input data consisted of: (i) mean weight at age (Table 11.2.2.2); (ii) maturity ogive (Table 11.2.2.3); (iii) natural mortality, $M$, that was set to 0.1 for all age groups in all years, except for 2009, where it was set 0.49 because of the Ichthyophonus infection, and for 2010 where M was for same reasons age dependent (Table 11.3.2.1; Óskarsson and Pálsson 2013); (iv) proportion of M before spawning was set to 0.5 ; and (v) proportion of F before spawning was set to 0 . Thus, no changes in the input data from last except for one more year of data.

## NFT-Adapt:

The estimated parameters in NFT Adapt are the stock in numbers-at-age. The parameters are output by the Levenburg-Marquardt Non-Linear Least Squares minimization algorithm (see VPA/ADAPT Version 3.3.0, Reference Manual). The estimated parameters were stock numbers for ages 4-12 in the end of year 2015, while the stock num-bers-at-age 3 were set to the geometric mean from 1991-2012. Like in last years' assessments, the input partial recruitment was set to 1 for ages 4 and older and the classic method was used to calculate the value of fully-recruited fishing mortality in the terminal year.

The catchability at age in the survey, as estimated by the NFT Adapt, and the CV is shown in Figure 11.3.2.1. The age groups 3-10 were used for tuning (Table 11.1.1.1 as decided at the benchmark in ICES (2011).

The output and model settings of the NFT-Adapt run (the adopted final assessment model; see below) are shown in Table 11.3.2.2. Stock numbers and fishing mortalities derived from the run are shown in Table 11.3.2.3 and Table 11.3.2.4, respectively, and summarized in Table 11.3.2.5 and Figure 11.3.2.2.

Residuals of the model fit are shown in Figure 11.3.2.3 and Table 11.3.2.6, and shows both cohort and year affects. The main pattern is the same as presented in recent assessments. Positive residuals, where the model estimates is smaller than seen in the survey, can be seen for 1994 and 1999 year classes for almost all age groups and a negative residuals for the 2001 and 2003 year classes. Year blocks of positive residuals are apparent for the years 2000-2006 (i.e. referring to January 1st), indicating that the model estimated the age groups smaller than observed in the surveys. During these years, the stock was overwintering in offshore areas off the east and west coast, compare to mainly easterly distribution before and overwintering in inshore areas there after (from 2006-2012). These positive blocks could therefore reflect changes in catchability of the survey for these years. Positive residuals, even if relatively weaker, were also observed for 2012. A block of negative residuals was however observed for 2009 (survey in the autumn 2008).

Retrospective analysis (Figure 11.3.2.4) indicate a more stability for the most recent three years than often before, i.e. adding new data to the model does not change the present perception of the stock size. The same applies correspondingly to the fishing mortality. Furthermore, to sustain the high M in the input data for 2009 and 2010 because of the infection, SSB of the most recent four years lifts compared with the preceding years. It required also an increase in recruitment estimates as apparent on the retrospective plots of number-at-age 3. A revision of the number-at-age 3 of the 2008 and 2009 year classes (in 2011 and 2012) is also apparent retrospectively, which is related to their high survey indices at age 3. Note that the high F in 2012 (Figure 11.3.2.4)
is due to the mass mortality, which was added to the catches that year in the assessment as presented earlier (ICES 2014).
The estimated number-at-age 3 in the final year (2016) from the NFT-Adapt of 649 million (geometric mean) was replaced with 464 million, which derives from a projection from number-at-age 1 in the 2014 acoustic survey as described by Gudmundsdóttir et al. (2007) and recommended to use for the stock (ICES 2011a).
Like described before (ICES 2014), the main difference between observed and predicted survey values from the NFT-Adapt model was for the period 1999-2004, where the observed values were well above the predicted (Figure 11.3.2.5), otherwise they fitted relatively well. Like seen in the residual plot (Figure 11.3.2.3), the observed value for the 2009 survey was lower than predicted and the vice versa for the 2012 survey (referring to the beginning of the year; Figure 11.3.2.5). The low survey value in 2009 is likely underestimate due to distribution of the stock that year in Breiðafjörður (Óskarsson et al. 2010), while the reason for the positive block during 2000-2004 is not fully known even if mainly caused by the large 1999 year class (ICES 2014) and possibly changes in the catchability of the survey as suggested above. However, an exploratory run in NFT-Adapt done in the 2011 assessment (ICES 2011b) where these years were excluded in the tuning, did not change the point estimate of the stock size in the latest year (January 1st 2011), implying that the terminal point estimates in the final run was not driven by this residual block.

## TASACS:

TASACS (Skagen and Skålevik 2009) is a toolbox with several assessment tools (VPA, a separable model and ISVPA) with a range of options for objective functions and for handling of the plus group. The VPA module was used for this stock, but it has been used in assessment of Norwegian spring-spawning herring since 2008 (ICES 2015b). The VPA module uses the catch numbers-at-age to reconstruct each year class backwards in time starting with the survivors at the end of the last year with catch data or the oldest true age. The survivor numbers are estimated by fitting the stock numbers-at-age to survey indices at age, calibrated with catchabilities.

TASACS model parameters that are used in the present configuration are:

- Stock numbers, $\mathrm{N}(\mathrm{a}, \mathrm{Y})$ and $\mathrm{N}(\mathrm{A}-1, \mathrm{y})$ : Survivor numbers at the end of each year class cohort.
- Survey indices I are modelled as $I=q^{*} N$, i.e. assuming a linear relation between survey index and stock number. Survey catchabilities are constant over time but dependent on age except for the ages older than 8.
- Natural mortalities $\mathrm{M}(\mathrm{a}, \mathrm{y})$. The same as used in the NFT run that is only increased mortality in 2009 and 2010, other wise 0.1 for all years and ages.
The oldest age $A$ is regarded as a plus group and modelled as a dynamic pool with mortality equal to that of the oldest true age: $\mathrm{N}(\mathrm{A}, \mathrm{y}+1)=\mathrm{N}(\mathrm{A}, \mathrm{y})^{*} \exp (-\mathrm{M}(\mathrm{A}, \mathrm{y})-\mathrm{F}(\mathrm{A}-1, \mathrm{y}))$ $+\mathrm{N}(\mathrm{A}-1, \mathrm{y})^{*} \exp (-\mathrm{M}(\mathrm{A}-1, \mathrm{y})-\mathrm{F}(\mathrm{A}-1, \mathrm{y}))$. The plus group is not included in the likelihood function. The likelihood function is a sum of squared log residuals. In the present configuration, they are not weighted according to their variance.
Different from the NFT-Adapt updated assessment run where the input data ranged from the 1987/88-2015/16, the data used in TASACS began in 1975/76, the year starts on 1st July (1st January in NFT), and ages 2-13+ from catches and surveys are used (3-

13+ in NFT). The survey takes place 1-January in NFT, but in December in TASACS and proportion mature and the fishery before the survey is set as 0.42 .

## Comparisons of different models and to runs from previous year:

The results of the assessments made in NFT-Adapt and TASACS were very similar (Figure 11.3.2.6). SSB follows the same trend and is at very similar level throughout almost the whole time-series. The main difference is in the years with high M (20092010) and is caused by different assumption of timing of the survey in the year, i.e. NFT-Adapt assume it is at the beginning of the assessment year while TASACS a half year earlier, but the consequence of this high $M$ becomes similar as seen in 2011. SSB in the final year (2016) was 318 kt in NFT-Adapt and 337 kt in TASACS and the difference is related to that number-at-age 3 in the final year in NFT-Adapt of 464 millions was projected from the survey index from 2014 (see above) while TASACS used unrealistic high and poorly determined estimated number of 770 millions from number-atage 2 in the survey (Figure 11.3.2.6).

The results of the final NFT run in 2016 were in a good agreement to the run in 2015 (Figure 11.3.2.2). The main difference is related to the number-at-age 3 in 2015 (Figure 11.3.2.2c), which was based on prediction from survey estimation of number-at-age 1 in the 2015 assessment while estimated by NFT in the 2016 assessment.

### 11.3.3 Final assessment

This is an update assessment so the results of the NFT-Adapt were adopted as point estimator for the prediction and thus the basis for the advice as in recent years. The model settings and outputs are shown in Table 11.3.2.2 to Table 11.3.2.4 and Figure 11.3.2.2.

The assessment (Table 11.3.2.5 and Figure 11.3.2.2) indicates that the fishing mortality (weighed average for age $5-10$ ) was 0.22 in 2015 or at $\mathrm{F}_{\mathrm{pa}} \mathrm{F}_{\mathrm{ms}}=0.22$, which is the target. The low F during 2009 to 2011 was related to cautious TAC and apparently overestimation of mortality induced by the Ichthyophonus outburst. Notice that the estimated number of herring that died in Kolgrafafjörður in the two incidents of the mass mortalities (Óskarsson et al. 2013) were added to the catches in 2012 and is also included in the high F that year (Table 11.3.2.5 and Figure 11.3.2.2). The F related only to landings in 2012 came to approximately 0.22 .

### 11.4 Reference points

Precautionary approach reference points:
The Working Group has pointed out that managing this stock at an exploitation rate at or above $\mathrm{F}_{0.1}=\mathrm{F}_{\text {MSY }}=0.22$ has been successful in the past, despite biased assessments. The Study Group on Precautionary Reference Points for Advice on Fishery Management met in February 2003 and concluded that it was not considered relevant to change the $B_{l i m}$ from 200000 t . At this year's NWWG meeting, the PA reference points were verified and revised (Óskarsson and Guðmundsdóttir 2016). On basis of the stock-recruitment relationship deriving from time-series ranging from 1947-2015, keeping Blim=200 kt was considered reasonable as the Study Group on Precautionary Reference Points for Advice on Fishery Management concluded also in February 2003. Other PA reference points were derived from $\mathrm{B}_{\mathrm{lim}}$ and these data in accordance to the ICES Advice Technical Guidelines and became these: $\mathrm{B}_{\mathrm{pa}}=273 \mathrm{kt}\left(\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\lim } \times \mathrm{e}^{1.645 \sigma}\right.$, where $\left.\sigma=0.19\right)$; $\mathrm{F}_{\lim }=0.61$ ( F that leads to $\mathrm{SSB}=\mathrm{B}_{\lim ,}$ given mean recruitment); $\mathrm{F}_{\mathrm{pa}}=0.43$ ( $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\lim } \times$ $\exp (-1.645 \times \sigma)$, where $\sigma=0.18)$.

In relation to the reference points it should be noted that the fishing mortality during 1987 to 2008 was on average 0.31 (weighed $\mathrm{F}_{5-10}$ ), or approximately $40 \%$ higher than the intended target of $\mathrm{F}_{0.1}=0.22$, but below the revised $\mathrm{F}_{\mathrm{pa}}$. This high F was despite the fact that the managers followed the scientific advice and restricted quotas with the aim of fishing at the intended target. Nevertheless, during this period, SSB remained above $B_{\text {lim }}$ and reached a record high level around 2008.

## MSY based reference points:

The MSY based reference points have not been set for Icelandic summer-spawning herring, but exploratory work was present at the NWWG meeting in 2011 in a form as requested by ICES (ICES 2011b). The HCS program Version 10.3 (Skagen, 2012) was used to evaluate possible points based on the MSY framework that could be a basis for a management plan and Harvest Control Rule later.

Number of different runs was made with varying settings. The results implied that the MSY framework was confirmative with the currently used precautionary reference points. It means that the currently used $\mathrm{F}_{0.1}=0.22$ could be a valid candidate for $\mathrm{F}_{\text {msy. }}$. This however, needs to be explored more thoroughly later. Development and testing of harvest control rule for the stock is also planned in 2016/2017

### 11.5 State of the stock

The stock was at high levels until 2008 but since then a substantial reduction took place despite a low fishing mortality. The reduction was caused by mortality induced by Ichthyophonus infection in the stock in 2008 and 2009. However, the observed high prevalence of infection for all the years since then is not considered to be causing further mortality in the stock. The continuing negative trend in the stock size since then, even if SSB is still above BPA, is due to a simultaneously negative trend in size of incoming of year classes, which have been below average size since appearance of the 2007 year class.

### 11.6 Short-term forecast

### 11.6.1 The input data

The final adopted model, NFT-Adapt, which gave the number-at-age on January $1^{\text {st }}$, 2016, was used for the prognosis. All input values for the prognosis are given in Table 11.6.1.1. Like done in the 2015 assessment for the first time, the number-at-age 3 in the assessment year (2013 year class) was set as the predicted number from juvenile survey (detailed below) instead of using geometric mean.

The weights were estimated from the last year catch weights (see Stock Annex) and as in the recent years, the weights are expected to continue to be high even if bit lower than predicted in last year's assessment (Figure 11.6.1.1). The selection pattern used in the prognosis was based on averages over 2013 to 2015 from the final run (Figure 11.6.1.2) (see Stock Annex). As traditionally, M was set 0.1 , proportion M before spawning was set 0.5 and proportion $F$ before spawning was set 0 . The numbers of recruits in the prognosis were determined as follows:

The 2013 year class: An acoustic survey aimed for getting an abundance index for this year class took place in November 2014 (Óskarsson and Reynisson 2015), and using a relation obtained by Gudmundsdóttir et al. (2007) provides estimate of 464 million at age 3 in 2016.

The 2014 year class: An acoustic survey aimed for getting an abundance index for this year class took place in September-October 2015 (Óskarsson 2016), and using a relation obtained by Gudmundsdóttir et al. (2007) provides estimate of 391 million at age 3 in 2017.

The 2015 year class: No acoustic estimates are available for the year class yet thus the number-at-age 3 in 2018 was set to the geometrical mean for age- 3 over 1987-2012, which give 596 million.

### 11.6.2 Prognosis results

SSB and biomass of age $3+$ are estimated to be 318 kt and 393 kt , respectively, at the beginning of the fishing season 2016/17 (approximately the same as at spawning in July 2016). The results of the short-term prediction from the final NFT-Adapt run (Table 11.6.1.2) indicate that fishing at 0.22 ( $=\mathrm{F}_{0.1}$; the stock is managed at $\mathrm{F}=0.22 \sim \mathrm{~F}_{\mathrm{MSY}}$ ) would correspond to TAC in 2016/2017 of 63 kt and SSB at the spawning season in 2016 would be 303 kt , or just above $\mathrm{B}_{\mathrm{pa}}$ of 300 kt .

The proposed composition of the catch in the season 2016/17 consists mainly of the 2008, 2009 and 2010 year classes, each contributing to $16-19 \%$ in total biomass of the catch (Figure 11.6.2.1). The small 2011 year class, at age 5 , will only give $8 \%$ of the catches.

### 11.7 Medium term predictions

Prognosis was made for the stock until the spawning season 2019 (Table 11.6.1.3) and the input data were the same as introduced above in section 11.6.1. The main features are that fishing at target $\mathrm{F}=0.22$ will give relatively constant catches and the SSB will remain above $B_{p a}$ and increase slightly throughout the period.

### 11.8 Uncertainties in assessment and forecast

### 11.8.1 Assessment

There are number of factors that could lead to uncertainty in the assessment. Three of them are addressed here. As done in the recent three assessments, additional natural mortality caused by the Ichthyophonus infection is only set for the first two years instead of all years since 2009. While this approach is considered to reduce the uncertainty in the assessment, quantification of the infection mortality needs to be improved in the future, and is ongoing currently. However, it should be noted that changing M for 2009 and 2010 changes the historical perception of the stocks size but has insignificant impacts on the assessment of the final year and the resulting advice (ICES 2014).

The apparent new infection in the stock in last two years is not considered to cause induced natural mortality in the stock and the fixed M of 0.1 is applied for those years. This decision was taken on basis of studies of the infection outbreak since 2009. The possibility of increased $M$ for these two years can however not be fully rejected and must be considered to add uncertainty to the assessment.

The part of the acoustic survey estimate deriving from east of Iceland in the autumn 2015 and represented $14 \%$ of the total biomass index, was considered to be with higher uncertainty, even if not estimated, than normally and elsewhere (see above and in Óskarsson 2016). This produces a small increase in uncertainty, even if unquantifiable, of the analytical assessment of the stock.

### 11.8.2 Forecast

The uncertainty in the assessment mentioned above related to the apparent new infection in the stock in last two years and the survey uncertainty applies also for the forecast.

The number-at-age 3 at the beginning of 2016 used in the prognosis ( 464 millions) was predicted from a survey estimate of number-at-age 1 in 2014 in accordance with the approach described in the Stock Annex. The size of the year class is therefore poorly determined and creates some uncertainty in the forecast.

### 11.8.3 Assessment quality

In previous years there has been concerns regarding the assessment because of retrospective patterns of the models. No assessment was provided in the 2005 due to data and model problems and in the two next consecutive years, ACFM rejected the assessment due to the retrospective pattern. In the assessments in 2007-2009 there was observed an improvement in the pattern from NFT-Adapt, while in 2010-2011, a retrospective pattern appeared again which was both related to the high M because of the Ichthyophonus infection but also due to new and more optimistic information about incoming year classes to the fishable stock (particularly the 2008 year class) and fishing pattern in recent year. The retrospective pattern in the last three and this year's assessment are less than seen for many years for SSB and F. That could be interpreted as an indication for improvements in the assessment quality compared with recent years.

### 11.9 Comparison with previous assessment and forecast

This year's assessment was conducted in the same way as in last year. In the current assessment, SSB at the beginning of the year 2015 is $1 \%$ higher ( 346 kt vs. 342 kt ), size of the 2010 year class $13 \%$ higher, size of the 2011 year class $37 \%$ higher, and $\mathrm{WF}_{5-10}$ in 2014 is $4 \%$ higher ( 0.255 vs. 0.266 ), compare to the 2015 assessment. Thus the assessment results are in a good agreement.

### 11.10 Management plans and evaluations

The practice has been to manage fisheries on this stock at $\mathrm{F}=\mathrm{F}_{0.1}\left(=0.22=\mathrm{F}_{\mathrm{pa}}\right)$ for more than 20 years. However, no formal management strategy has been developed and proposed but such work is planned to be initiated in 2016.

### 11.11 Management consideration

Inspections indicate still a high prevalence of heart lesions related to Ichthyophonus hoferi in the herring stock, as in the last seven years. Last winter, and to some degree the winter before, new infection has been taken place in the stock, which has not been seen since 2011. Induced mortality due to the infection is considered insignificant for the years 2011-2016 in the assessment, but for the most recent two years such an induced mortality cannot be fully rejected yet and must be considered to add uncertainty to the assessment.

### 11.12 Ecosystem considerations

The reason for the outbreak of Ichthyophonus infection in the herring stock that was first observed in the autumn 2008 is not known but is probably the effect of interaction between environmental factors and distribution of the stock (Óskarsson et al. 2009). It includes that outbreak of Ichthyophonus spores in the environment, which infect the
herring via oral intake (Jones and Dawe 2002), could be linked to the observed increased temperature off the southwest coast. Further researches on the causes of such an outbreak are needed and how the herring get infected, i.e. through intake of free floating spores or through zooplankton that contain spores. However, with respect of the impacts of the outbreak on the herring stock, significant additional mortality was estimated to have taken place only in the first two years (ICES 2014; Óskarsson and Pálsson 2013), despite a high prevalence of infection for now eight years. Thus, the infection that is still found in the stock (Óskarsson and Pálsson 2016) will most likely decrease and disappear over some years as the fish gets older. The observed new infection in the winter 2015/16 can however delay this process.

The WG does not have any information of direct evidence of environmental effects of the stock but emphasize that increased sea temperature is considered to have generally positive effects on the stock (Jakobsson and Stefansson, 1999; Óskarsson and Taggart 2010). It is manifest in observations of larger number of recruits per SSB during warm years and relatively high mean weight-at-age during recent years. Furthermore, the stock occupies colder water around Iceland than other herring stocks in the N -Atlantic and is therefore on edge of the distribution towards cold water, where warming will generally have a positive impacts on the stock development. The increased temperature in Icelandic waters since 1998 (MRI 2012), has therefore probably positive effects on the stock, possibly apart from the Ichthyophonus outbreak.

### 11.13 Regulations and their effects

The fishery of the Icelandic summer-spawning herring is limited to the period 1 September to 1 May each season, according to regulations set by the Icelandic Fishery Ministry (no. 770, 8. September 2006). Several other regulations are enforced by the Ministry that effect the herring fishery. They involve protections of juveniles herring ( 27 cm and smaller) in the fishery where area closures are enforced if the proportion of juveniles exceeds $25 \%$ in number (no. 376, 8 . October 1992). No such closures took place in $2015 / 2016$. Another regulation deals with the quantity of bycatch allowed. Then there are regulation that prohibit use of pelagic trawls within the 12 nm fishing zone (no. 770, 8. September 2006), which is enforced to limit bycatch of juveniles of other fish species.

### 11.14 Changes in fishing technology and fishing patterns

There are no recent changes in fishing technology which may lead to different catch compositions. The fishing pattern in 2014/2015 and 2015/2016 was different from the previous seven seasons. Instead of fishing near only in a small inshore area off the west coast in purse-seine, the whole directed fishery took place in offshore areas west of Iceland by pelagic trawls. These changes are not considered to affect the selectivity of the fishery because the fishery is still targeting dense schools of overwintering herring in large fishing gears, getting huge catches in each haul and is by none means size selective.

Bycatch of Icelandic summer-spawning herring in summer fishery for NE-Atlantic mackerel and Norwegian spring-spawning herring has been taken place since around mid-2000s. Until that time, no summer fishery on this stock had taken place for decades. This bycatch of summer spawners is partly on the stock components (e.g. juveniles and herring east of Iceland) that are not fished in the direct fishery on the
overwintering grounds in the west. However, this bycatches are well sampled and contributes to less than $10 \%$ of the total annual catch (except for $13 \%$ in 2014/2015) so the impacts of these changes on the assessment are considered to be insignificant.
The fishing pattern varies annually as noted in section 11.2 and it is related to variation in distribution of the different age classes of the stock. This variation can have consequences for the catch composition but it is impossible to provide a forecast about this variation.

### 11.15 Species interaction effects and ecosystem drivers

The WG have not dealt with this issue in a thoroughly and dedicated manner. However, some work has been done in this field in recent years in one way or another.

Regarding relevant researches on species interaction, the main work relates to the increasing amount of Northeast Atlantic mackerel (NEAM) feeding in Icelandic waters since 2007 (Astthorsson et al. 2012; Nøttestad et al. 2016). Surveys in the summers since 2010 indicate a high overlap in spatial and temporal distribution of NEAM and Icelandic summer-spawning herring (Óskarsson et al. 2016). Moreover, the diet composition of NEAM in Icelandic waters showed a clear overlap with those of the two herring stocks, i.e. Icelandic summer-spawning herring and Norwegian spring-spawning herring (Óskarsson et al. 2016). Even if Copepoda was important diet group for all the three stocks its relative contribution to the total diet was apparently higher for NEAM than the two herring stocks. Considering former studies of herring diet, this finding was unexpected, and particularly how little the Copepoda contributed to the herring diet. This difference in the stomach content of NEAM and the two herring stocks indicated that there could be some difference in feeding ecology between them in Icelandic waters, where NEAM preferred Copepoda, or feed in the water column where they dominate over other prey groups, while the opposite would be for the herring and the prey Euphausiacea. Recent studies in the Nordic Seas have shown similar results (Langøy et al. 2012; Debes et al. 2012). The indication for difference in feeding ecology of the species is further supported by the fact that the body condition of the two herring stocks showed no clear decreasing trend since the invasion of NEAM started into Icelandic waters. On the contrary the mean weights-at-age of the summer spawners have been high, for example record high in the autumn 2014 (Figure 11.6.1.1), and the mean weight-at-length have also been relatively high in recent years (Óskarsson and Pálsson 2015). It should though be noted that comparison of the diet composition of herring in recent years to earlier studies, mainly on NSS herring, indicate that the herring might have shifted their feeding preference towards Euphausiacea instead of Copepoda. That is possibly a consequence of increased competition for food with NEAM, where the herring is overwhelmed and shifts towards other preys.

The WG is not aware of documentations of strong signals from ecosystem or environmental variables that impact the herring stock and could possibly be a basis for implementing ecosystem drivers in the analytical basis for its advice. For example, recruitment in the stock has been positively, but weakly, linked to NAO winter index (North Atlantic Oscillation) and sea temperature (Óskarsson and Taggart 2010), while indices representing zooplankton abundance in the spring have not been found to impact the recruitment (Óskarsson and Taggart 2010) or body condition and growth rate of the adult part of the stock (Óskarsson 2008).

### 11.16 Comments on the PA reference points

The WG dealt with reference points at this year's meeting as mentioned in section 11.4 and have revised them in accordance to the ICES Advice Technical Guidelines and they are considered reasonable. The analyses are detailed in Óskarsson and Gudmundsdottir (2016).

### 11.17 Comments on the assessment

The assessment implies that the stock size has been rather unvarying in recent years following a period of depletion related to the Ichthyophonus infection. This is related to average size recruiting year classes entering the fishable stock, no infection mortality, and moderate fishing mortality. The assessment follows fairly well the pattern in the tuning series for recent years (Figure 11.3.2.5). However, small year class from 20092012, particularly 2011, entering the spawning stock causes a decline in SSB even if it still above $B_{p a}$. The size of the small 2011 year class is less pessimistic in this assessment but the downward trend is foreseen anyway.

This year's research on the Ichthyophonus infection in the stock supports the approach taken since 2013 (ICES 2013; Óskarsson and Pálsson 2013) that additional natural mortality in the stock due to Ichthyophonus infection should only be applied for the first two years of the outburst. Further research is ongoing to quantify the infection mortality for these two years, but it is important to note that changing the mortality in the assessment has mainly impacts on the historical perspective of the stock size and insignificant impacts on the present stock status.
In conclusion of the review group for NWWG 2011 (ICES 2011b), the suggestion was "to improve the assessment in order to get a better fitting for the years 2000-2005 and to work on the reference points". In this year's assessment, the reference points issue was in part revisited (see above), while the other issue was not dealt with. The years 2000-2005 fit still poorly to the tuning series and no satisfactory explanation exists for this pattern. The models recently used for the stock (NFT-Adapt, TASACS, TSA and Coleraine; ICES 2011a) are not able to follow this trend in the tuning series. It should be noted that this same pattern was observed in the benchmark assessment in 2011 (Gudmundsdottir 2011) where input data were limited to the period before the infection so assumptions related to the natural mortality-infection are probably only responsible for this pattern to small degree if any. As mention above (section 11.3.2), the discrepancy could be related to the fact that during these years, the stock was overwintering in offshore areas off the east and west coast, compare to mainly easterly distribution before and overwintering in inshore areas there after (from 2006-2013). These positive blocks could therefore reflect changes in catchability of the survey for these years. This must be kept in mind for the years to come since the stock has now started again to overwinter in offshore areas.

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### 11.19 Tables

Table 11.1.1.1. Icelandic summer-spawning herring. Acoustic estimates (in millions) in the winters 1973/74-2015/16 (age refers to the autumns). No surveys (and gaps in the time-series) were in 1976/77, 1982/83, 1986/87, 1994/95.

| Year \age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1973/74 | 154.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 154 |
| 1974/75 | 5.000 | 137.000 | 19.000 | 21.000 | 2.000 | 2.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 186 |
| 1975/76 | 136.000 | 20.000 | 133.000 | 17.000 | 10.000 | 3.000 | 3.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 322 |
| 1977/78 | 212.000 | 424.000 | 46.000 | 19.000 | 139.000 | 18.000 | 18.000 | 10.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 886 |
| 1978/79 | 158.000 | 334.000 | 215.000 | 49.000 | 20.000 | 111.000 | 30.000 | 30.000 | 20.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 967 |
| 1979/80 | 19.000 | 177.000 | 360.000 | 253.000 | 51.000 | 41.000 | 93.000 | 10.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1004 |
| 1980/81 | 361.000 | 462.000 | 85.000 | 170.000 | 182.000 | 33.000 | 29.000 | 58.000 | 10.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1390 |
| 1981/82 | 17.000 | 75.000 | 159.000 | 42.000 | 123.000 | 162.000 | 24.000 | 8.000 | 46.000 | 10.000 | 0.000 | 0.000 | 0.000 | 0.000 | 666 |
| 1983/84 | 171.000 | 310.000 | 724.000 | 80.000 | 39.000 | 15.000 | 27.000 | 26.000 | 10.000 | 5.000 | 12.000 | 0.000 | 0.000 | 0.000 | 1419 |
| 1984/85 | 28.000 | 67.000 | 56.000 | 360.000 | 65.000 | 32.000 | 16.000 | 17.000 | 18.000 | 9.000 | 7.000 | 4.000 | 5.000 | 5.000 | 689 |
| 1985/86 | 652.000 | 208.000 | 110.000 | 86.000 | 425.000 | 67.000 | 41.000 | 17.000 | 27.000 | 26.000 | 16.000 | 6.000 | 6.000 | 1.000 | 1688 |
| 1987/88 | 115.544 | 401.246 | 858.012 | 308.065 | 57.103 | 32.532 | 70.426 | 36.713 | 23.586 | 18.401 | 24.278 | 10.127 | 3.926 | 4.858 | 1965 |
| 1988/89 | 635.675 | 201.284 | 232.808 | 381.417 | 188.456 | 46.448 | 25.798 | 32.819 | 17.439 | 10.373 | 9.081 | 5.419 | 3.128 | 5.007 | 1795 |
| 1989/90 | 138.780 | 655.361 | 179.364 | 278.836 | 592.982 | 179.665 | 22.182 | 21.768 | 13.080 | 9.941 | 1.989 | 0.000 | 0.000 | 0.000 | 2094 |
| 1990/91 | 403.661 | 132.235 | 258.591 | 94.373 | 191.054 | 514.403 | 79.353 | 37.618 | 9.394 | 12.636 | 0.000 | 0.000 | 0.000 | 0.000 | 1733 |
| 1991/92 | 598.157 | 1049.990 | 354.521 | 319.866 | 89.825 | 138.333 | 256.921 | 21.290 | 9.866 | 0.000 | 9.327 | 0.000 | 0.000 | 1.494 | 2850 |
| 1992/93 | 267.862 | 830.608 | 729.556 | 158.778 | 130.781 | 54.156 | 96.330 | 96.649 | 24.542 | 1.130 | 1.130 | 3.390 | 0.000 | 0.000 | 2395 |
| 1993/94 | 302.075 | 505.279 | 882.868 | 496.297 | 66.963 | 58.295 | 106.172 | 48.874 | 36.201 | 0.000 | 4.224 | 18.080 | 0.000 | 0.000 | 2525 |
| 1995/96 | 216.991 | 133.810 | 761.581 | 277.893 | 385.027 | 176.906 | 98.150 | 48.503 | 16.226 | 29.390 | 47.945 | 4.476 | 0.000 | 0.000 | 2197 |
| 1996/97 | 33.363 | 270.706 | 133.667 | 468.678 | 269.888 | 325.664 | 217.421 | 92.979 | 55.494 | 39.048 | 30.028 | 53.216 | 18.838 | 12.612 | 2022 |
| 1997/98 | 291.884 | 601.783 | 81.055 | 57.366 | 287.046 | 155.998 | 203.382 | 105.730 | 35.469 | 27.373 | 14.234 | 36.500 | 14.235 | 11.570 | 1924 |
| 1998/99 | 100.426 | 255.937 | 1081.504 | 103.344 | 51.786 | 135.246 | 70.514 | 101.626 | 53.935 | 17.414 | 13.636 | 2.642 | 4.209 | 8.775 | 2001 |
| 1999/00 | 516.153 | 839.491 | 239.064 | 605.858 | 88.214 | 43.353 | 165.716 | 89.916 | 121.345 | 77.600 | 21.542 | 3.740 | 11.149 | 0.000 | 2823 |


| Year \age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000/01 | 190.281 | 966.960 | 1316.413 | 191.001 | 482.418 | 34.377 | 15.727 | 37.940 | 14.320 | 15.413 | 14.668 | 1.705 | 3.259 | 0.000 | 3284 |
| 2001/02 | 1047.643 | 287.004 | 217.441 | 260.497 | 161.049 | 345.852 | 62.451 | 57.105 | 38.405 | 46.044 | 38.114 | 21.062 | 3.663 | 0.000 | 2586 |
| 2002/03 | 1731.809 | 1919.368 | 553.149 | 205.656 | 262.362 | 153.037 | 276.199 | 99.206 | 47.621 | 55.126 | 18.798 | 24.419 | 24.112 | 1.377 | 5372 |
| 2003/04 | 1115.255 | 1434.976 | 2058.222 | 330.800 | 109.146 | 100.785 | 38.693 | 45.582 | 7.039 | 6.362 | 7.509 | 10.894 | 0.000 | 2.289 | 5268 |
| 2004/05 | 2417.128 | 713.730 | 1022.326 | 1046.657 | 171.326 | 62.429 | 44.313 | 10.947 | 23.942 | 12.669 | 0.000 | 1.948 | 11.088 | 0.000 | 5539 |
| 2005/06 | 469.532 | 443.877 | 344.983 | 818.738 | 1220.902 | 281.448 | 122.183 | 129.588 | 73.339 | 65.287 | 10.115 | 9.205 | 3.548 | 12.417 | 4005 |
| 2006/07 | 109.959 | 608.205 | 1059.597 | 410.145 | 424.525 | 693.423 | 95.997 | 123.748 | 48.773 | 0.955 | 0.000 | 0.000 | 0.000 | 0.480 | 3576 |
| 2007/08 | 90.231 | 456.773 | 289.260 | 541.585 | 309.443 | 402.889 | 702.708 | 221.626 | 244.772 | 13.997 | 22.113 | 68.105 | 10.136 | 2.800 | 3376 |
| 2008/09 | 149.466 | 196.127 | 416.862 | 288.156 | 457.659 | 266.975 | 225.747 | 168.960 | 29.922 | 26.281 | 17.790 | 9.881 | 0.974 | 3.195 | 2258 |
| 2009/10 | 151.066 | 315.941 | 490.653 | 554.818 | 271.445 | 327.275 | 149.143 | 83.875 | 156.920 | 36.666 | 13.649 | 8.507 | 1.458 | 5.590 | 2567 |
| 2010/11 | 106.178 | 280.582 | 228.857 | 304.885 | 296.254 | 138.686 | 301.285 | 60.997 | 141.323 | 97.412 | 37.006 | 0.000 | 4.019 | 0.000 | 1997 |
| 2011/12 | 704.863 | 977.323 | 434.876 | 313.742 | 272.140 | 239.320 | 154.581 | 175.088 | 84.582 | 92.435 | 89.376 | 17.638 | 6.808 | 4,989 | 3676 |
| 2012/13 | 178.500 | 781.083 | 631.421 | 166.627 | 126.961 | 142.044 | 110.084 | 97.000 | 74.340 | 69.473 | 43.376 | 38.450 | 7.458 | 0.773 | 2468 |
| 2013/14 | 15.919 | 314.865 | 218.715 | 344.981 | 151.631 | 132.767 | 120.756 | 118.377 | 89.555 | 74.602 | 48.695 | 44.637 | 31.096 | 11.598 | 1718 |
| 2014/15 | 152.422 | 90,269 | 330.084 | 260.919 | 259.079 | 187.905 | 111.955 | 91.629 | 37.855 | 76.680 | 30.366 | 10.619 | 22.799 | 10.108 | 1667 |
| 2015/16 | 381.900 | 164.221 | 174.507 | 312.350 | 225.836 | 215.207 | 93.743 | 62.753 | 75.339 | 41.961 | 15.696 | 26.756 | 20.159 | 5.401 | 1816 |

Table 11.1.1.2. Icelandic summers-spawning herring. Number of scales by ages and number of samples taken in the annual acoustic surveys in the seasons $1987 / 88$ 2015/16 (age refers to the former year, i.e. autumns). In 2000 seven samples were used from the fishery. No survey was conducted in 1994/95.

| Year\AGE | Number of scales |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Number of samples |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | Total | Total | West | EAST |
| 1987/88 | 11 | 59 | 246 | 156 | 37 | 28 | 58 | 33 | 22 | 16 | 23 | 10 | 5 | 8 | 712 | 8 | 1 | 7 |
| 1988/89 | 229 | 78 | 181 | 424 | 178 | 69 | 50 | 77 | 42 | 29 | 23 | 13 | 7 | 12 | 1412 | 18 | 5 | 10 |
| 1989/90 | 38 | 245 | 96 | 132 | 225 | 35 | 2 | 2 | 3 | 3 | 2 | 0 | 0 | 0 | 783 | 8 |  | 8 |
| 1990/91 | 418 | 229 | 303 | 90 | 131 | 257 | 28 | 6 | 3 | 8 | 0 | 0 | 0 | 0 | 1473 | 15 |  | 15 |
| 1991/92 | 414 | 439 | 127 | 127 | 33 | 48 | 84 | 5 | 3 | 0 | 2 | 0 | 0 | 1 | 1283 | 15 |  | 15 |
| 1992/93 | 122 | 513 | 289 | 68 | 73 | 28 | 38 | 34 | 6 | 2 | 2 | 6 | 0 | 0 | 1181 | 12 |  | 12 |
| 1993/94 | 63 | 285 | 343 | 129 | 13 | 15 | 7 | 14 | 11 | 0 | 1 | 3 | 0 | 0 | 884 | 9 |  | 9 |
| 1994/95* |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1995/96 | 183 | 90 | 471 | 162 | 209 | 107 | 38 | 18 | 8 | 14 | 18 | 2 | 0 | 0 | 1320 | 14 | 9 | 5 |
| 1996/97 | 24 | 150 | 88 | 351 | 141 | 137 | 87 | 32 | 15 | 10 | 7 | 14 | 4 | 2 | 1062 | 11 | 4 | 7 |
| 1997/98 | 101 | 249 | 50 | 36 | 159 | 95 | 122 | 62 | 21 | 13 | 8 | 15 | 8 | 5 | 944 | 14 | 7 | 7 |
| 1998/99 | 130 | 216 | 777 | 72 | 31 | 65 | 59 | 86 | 37 | 22 | 17 | 5 | 6 | 11 | 1534 | 17 | 10 | 7 |
| 1999/00 | 116 | 227 | 72 | 144 | 17 | 13 | 26 | 26 | 27 | 10 | 8 | 2 | 1 | 0 | 689 | 7 | 3 | 4 |
| 2000/01 | 116 | 249 | 332 | 87 | 166 | 10 | 7 | 21 | 8 | 14 | 11 | 3 | 1 | 0 | 1025 | 14 | 10 | 4 |
| 2001/02 | 61 | 56 | 130 | 114 | 62 | 136 | 25 | 24 | 17 | 21 | 17 | 10 | 3 | 0 | 676 | 9 | 4 | 5 |
| 2002/03 | 520 | 705 | 258 | 104 | 130 | 74 | 128 | 46 | 26 | 25 | 13 | 15 | 10 | 1 | 2055 | 22 | 12 | 10 |
| 2003/04 | 126 | 301 | 415 | 88 | 35 | 32 | 15 | 17 | 3 | 4 | 4 | 6 | 1 | 1 | 1048 | 13 | 8 | 5 |
| 2004/05 | 304 | 159 | 284 | 326 | 70 | 29 | 17 | 5 | 8 | 4 | 0 | 3 | 3 | 0 | 1212 | 13 | 4 | 9 |
| 2005/06 | 217 | 312 | 190 | 420 | 501 | 110 | 40 | 38 | 26 | 18 | 5 | 5 | 5 | 7 | 1894 | 22 | 14 | 8 |
| 2006/07 | 19 | 77 | 134 | 64 | 71 | 88 | 22 | 4 | 2 | 2 | 0 | 0 | 0 | 1 | 484 | 6 | 4 | 2 |
| 2007/08 | 58 | 288 | 180 | 264 | 85 | 80 | 104 | 19 | 15 | 2 | 2 | 6 | 1 | 3 | 1107 | 17 | 13 | 4 |
| 2008/09 | 274 | 208 | 213 | 136 | 204 | 123 | 125 | 97 | 18 | 13 | 9 | 7 | 4 | 17 | 1448 | 29 | 19 | 10 |
| 2009/10 | 104 | 100 | 105 | 116 | 60 | 74 | 34 | 19 | 36 | 8 | 3 | 4 | 2 | 2 | 667 | 17 | 10 | 7 |
| 2010/11 | 35 | 74 | 102 | 157 | 139 | 61 | 119 | 22 | 52 | 36 | 13 | 0 | 1 | 0 | 811 | 11 | 8 | 3 |
| 2011/12 | 229 | 330 | 134 | 115 | 100 | 106 | 74 | 87 | 45 | 48 | 51 | 10 | 3 | 3 | 1335 | 15 | 9 | 6 |


| Year \age | Number of scales |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Number of Samples |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | Total | Total | West | EAST |
| $2012 / 13 \ddagger$ | 42 | $266$ | 554 | 273 | 220 | 252 | 198 | 165 | 126 | 114 | 69 | 61 | 12 | 2 | 2370 | 60 | $55 \ddagger$ | 5 |
| 2013/14 | 26 | 472 | 275 | 414 | 199 | 200 | 199 | 208 | 163 | 138 | 90 | 85 | 60 | 23 | 2552 | 45 | $37 \ddagger$ | 8 |
| 2014/15 | 83 | 50 | 96 | 71 | 72 | 53 | 32 | 26 | 11 | 22 | 8 | 3 | 6 | 4 | 534 | 10 | 8 | 2 |
| 2015/16 | 229 | 112 | 131 | 208 | 148 | 123 | 47 | 32 | 32 | 22 | 13 | 7 | 12 | 4 | 1120 | 14 | 7 | 7§ |

*No survey
$\ddagger$ Samples in the western part were mainly from the commercial catch as there was impossible to secure a usable research survey samples from Kolgrafafjörður where most of the herring was observed.
$\$ 3$ samples were taken in the east and south in this survey (B1-2016), while four were taken in the west and used also in the age-length key.

Table 11.2.1. Icelandic summer spawners. Landings, catches, recommended TACs, and set National TACs in thousand tonnes.

| Year | LANDINGS | Catches | Recom. <br> TACs | NAT. <br> TACs | Year | LANDINGS | Catches | Recom. <br> TACs | $\begin{aligned} & \text { NAT. } \\ & \text { TACs } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 0.31 | 0.31 |  |  | 2007/2008 | 158.9 | 158.9 | 130 | 150 |
| 1973 | 0.254 | 0.254 |  |  | 2008/2009 | 151.8 | 151.8 | 130 | 150 |
| 1974 | 1.275 | 1.275 |  |  | 2009/2010 | 46.3 | 46.3 | 40 | 47 |
| 1975 | 13.28 | 13.28 |  |  | 2010/2011 | 43.5 | 43.5 | 40 | 40 |
| 1976 | 17.168 | 17.168 |  |  | 2011/2012 | 49.4 | 49.4 | 40 | 45 |
| 1977 | 28.925 | 28.925 |  |  | 2012/2013 | 72.0 | 72.0 | 67 | 68.5 |
| 1978 | 37.333 | 37.333 |  |  | 2013/2014 | 72.0 | 72.0 | 87 | 87 |
| 1979 | 45.072 | 45.072 |  |  | 2014/2015§ | 95.0 | 95.0 | 83 | 83 |
| 1980 | 53.268 | 53.268 |  |  | 2015/2016 | 69.7 | 69.7 | 71 | 71 |
| 1981 | 39.544 | 39.544 |  |  |  |  |  |  |  |
| 1982 | 56.528 | 56.528 |  |  |  |  |  |  |  |
| 1983 | 58.867 | 58.867 |  |  |  |  |  |  |  |
| 1984 | 50.304 | 50.304 |  |  |  |  |  |  |  |
| 1985 | 49.368 | 49.368 | 50 | 50 |  |  |  |  |  |
| 1986 | 65.5 | 65.5 | 65 | 65 |  |  |  |  |  |
| 1987 | 75 | 75 | 70 | 73 |  |  |  |  |  |
| 1988 | 92.8 | 92.8 | 90 | 90 |  |  |  |  |  |
| 1989 | 97.3 | 101 | 90 | 90 |  |  |  |  |  |
| 1990/1991 | 101.6 | 105.1 | 80 | 110 |  |  |  |  |  |
| 1991/1992 | 98.5 | 109.5 | 80 | 110 |  |  |  |  |  |
| 1992/1993 | 106.7 | 108.5 | 90 | 110 |  |  |  |  |  |
| 1993/1994 | 101.5 | 102.7 | 90 | 100 |  |  |  |  |  |
| 1994/1995 | 132 | 134 | 120 | 120 |  |  |  |  |  |
| 1995/1996 | 125 | 125.9 | 110 | 110 |  |  |  |  |  |
| 1996/1997 | 95.9 | 95.9 | 100 | 100 |  |  |  |  |  |
| 1997/1998 | 64.7 | 64.7 | 100 | 100 |  |  |  |  |  |


| Year | Landings | CATChes | Recom. <br> TACs | NAT. <br> TACs | Nat. <br> TACs |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $1998 / 1999^{* *}$ | 87 | 87 | 90 | 70 | TEAR |

*Summer fishery in 2002 and 2003 included
** TAC was decided 70 thous. tonnes but because of transfers from the previous quota year the national TAC became 90 thous. tonnes.
$\ddagger$ Landings and catches include bycatch of Icelandic summer-spawning herring in the mackerel and NSS herring fishery during the preceding summer (i.e. from the fishing season before in June-August).
§ The landings and catches in 2014/2015 consist of transfer of 7 kt from the year before and 5 kt from the year to come, which explains the discrepancy to the TACs.

Table 11.2.2.1. Icelandic summer-spawning herring. Catch in numbers (millions) and total catch in weight (thous. tonnes) ( 1981 refers to season 1981/1982 etc.).

| Year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ | Catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 1.518 | 2.049 | 31.975 | 6.493 | 7.905 | 0.863 | 0.442 | 0.345 | 0.114 | 0.004 | 0.001 | 0.001 | 0.001 | 0.001 | 13.280 |
| 1976 | 0.614 | 9.848 | 3.908 | 34.144 | 7.009 | 5.481 | 1.045 | 0.438 | 0.296 | 0.134 | 0.092 | 0.001 | 0.001 | 0.001 | 17.168 |
| 1977 | 0.705 | 18.853 | 24.152 | 10.404 | 46.357 | 6.735 | 5.421 | 1.395 | 0.524 | 0.362 | 0.027 | 0.128 | 0.001 | 0.001 | 28.925 |
| 1978 | 2.634 | 22.551 | 50.995 | 13.846 | 8.738 | 39.492 | 7.253 | 6.354 | 1.616 | 0.926 | 0.4 | 0.017 | 0.025 | 0.051 | 37.333 |
| 1979 | 0.929 | 15.098 | 47.561 | 69.735 | 16.451 | 8.003 | 26.04 | 3.05 | 1.869 | 0.494 | 0.439 | 0.032 | 0.054 | 0.006 | 45.072 |
| $1980$ | 3.147 | 14.347 | 20.761 | 60.727 | 65.328 | 11.541 | 9.285 | 19.442 | 1.796 | 1.464 | 0.698 | 0.001 | 0.11 | 0.079 | 53.268 |
| 1981 | 2.283 | 4.629 | 16.771 | 12.126 | 36.871 | 41.917 | 7.299 | 4.863 | 13.416 | 1.032 | 0.884 | 0.760 | 0.101 | 0.062 | 39.544 |
| 1982 | 0.454 | 19.187 | 28.109 | 38.280 | 16.623 | 38.308 | 43.770 | 6.813 | 6.633 | 10.457 | 2.354 | 0.594 | 0.075 | 0.211 | 56.528 |
| 1983 | 1.475 | 22.499 | 151.718 | 30.285 | 21.599 | 8.667 | 14.065 | 13.713 | 3.728 | 2.381 | 3.436 | 0.554 | 0.100 | 0.003 | 58.867 |
| 1984 | 0.421 | 18.015 | 32.244 | 141.354 | 17.043 | 7.113 | 3.916 | 4.113 | 4.517 | 1.828 | 0.202 | 0.255 | 0.260 | 0.003 | 50.304 |
| 1985 | 0.112 | 12.872 | 24.659 | 21.656 | 85.210 | 11.903 | 5.740 | 2.336 | 4.363 | 4.053 | 2.773 | 0.975 | 0.480 | 0.581 | 49.368 |
| 1986 | 0.100 | 8.172 | 33.938 | 23.452 | 20.681 | 77.629 | 18.252 | 10.986 | 8.594 | 9.675 | 7.183 | 3.682 | 2.918 | 1.788 | 65.500 |
| 1987 | 0.029 | 3.144 | 44.590 | 60.285 | 20.622 | 19.751 | 46.240 | 15.232 | 13.963 | 10.179 | 13.216 | 6.224 | 4.723 | 2.280 | 75.439 |
| 1988 | 0.879 | 4.757 | 41.331 | 99.366 | 69.331 | 22.955 | 20.131 | 32.201 | 12.349 | 10.250 | 7.378 | 7.284 | 4.807 | 1.957 | 92.828 |
| 1989 | 3.974 | 22.628 | 26.649 | 77.824 | 188.654 | 43.114 | 8.116 | 5.897 | 7.292 | 4.780 | 3.449 | 1.410 | 0.844 | 0.348 | 101.000 |
| 1990 | 12.567 | 14.884 | 56.995 | 35.593 | 79.757 | 157.225 | 30.248 | 8.187 | 4.372 | 3.379 | 1.786 | 0.715 | 0.446 | 0.565 | 105.097 |
| 1991 | 37.085 | 88.683 | 49.081 | 86.292 | 34.793 | 55.228 | 110.132 | 10.079 | 4.155 | 2.735 | 2.003 | 0.519 | 0.339 | 0.416 | 109.489 |
| 1992 | 16.144 | 94.86 | 122.626 | 38.381 | 58.605 | 27.921 | 38.42 | 53.114 | 11.592 | 1.727 | 1.757 | 0.153 | 0.376 | 0.001 | 108.504 |
| 1993 | 2.467 | 51.153 | 177.78 | 92.68 | 20.791 | 28.56 | 13.313 | 19.617 | 15.266 | 4.254 | 0.797 | 0.254 | 0.001 | 0.001 | 102.741 |
| 1994 | 5.738 | 134.616 | 113.29 | 142.876 | 87.207 | 24.913 | 20.303 | 16.301 | 15.695 | 14.68 | 2.936 | 1.435 | 0.244 | 0.195 | 134.003 |
| 1995 | 4.555 | 20.991 | 137.232 | 86.864 | 109.14 | 76.78 | 21.361 | 15.225 | 8.541 | 9.617 | 7.034 | 2.291 | 0.621 | 0.235 | 125.851 |
| 1996 | 0.717 | 15.969 | 40.311 | 86.187 | 68.927 | 84.66 | 39.664 | 14.746 | 8.419 | 5.836 | 3.152 | 5.18 | 1.996 | 0.574 | 95.882 |
| 1997 | 2.008 | 39.24 | 30.141 | 26.307 | 36.738 | 33.705 | 31.022 | 22.277 | 8.531 | 3.383 | 1.141 | 10.296 | 0.947 | 2.524 | 64.682 |
| 1998 | 23.655 | 45.39 | 175.529 | 22.691 | 8.613 | 40.898 | 25.944 | 32.046 | 14.647 | 2.122 | 2.754 | 2.15 | 1.07 | 1.011 | 86.998 |
| 1999 | 5.306 | 56.315 | 54.779 | 140.913 | 16.093 | 13.506 | 31.467 | 19.845 | 22.031 | 12.609 | 2.673 | 2.746 | 1.416 | 2.514 | 92.896 |
| 2000 | 17.286 | 57.282 | 136.278 | 49.289 | 76.614 | 11.546 | 8.294 | 16.367 | 9.874 | 11.332 | 6.744 | 2.975 | 1.539 | 1.104 | 100.332 |
| 2001 | 27.486 | 42.304 | 86.422 | 93.597 | 30.336 | 54.491 | 10.375 | 8.762 | 12.244 | 9.907 | 8.259 | 6.088 | 1.491 | 1.259 | 95.675 |


| YEAR\AGE | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5 +}$ | CATCH |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 11.698 | 80.863 | 70.801 | 45.607 | 54.202 | 21.211 | 42.199 | 9.888 | 4.707 | 6.52 | 9.108 | 9.355 | 3.994 | 5.697 | 96.128 |
| 2003 | 24.477 | 211.495 | 286.017 | 58.120 | 27.979 | 25.592 | 14.203 | 10.944 | 2.230 | 3.424 | 4.225 | 2.562 | 1.575 | 1.370 | 130.741 |
| 2004 | 23.144 | 63.355 | 139.543 | 182.45 | 40.489 | 13.727 | 9.342 | 5.769 | 7.021 | 3.136 | 1.861 | 3.871 | 0.994 | 1.855 | 114.237 |
| 2005 | 6.088 | 26.091 | 42.116 | 117.91 | 133.437 | 27.565 | 12.074 | 9.203 | 5.172 | 5.116 | 1.045 | 1.706 | 2.11 | 0.757 | 103.043 |
| 2006 | 52.567 | 118.526 | 217.672 | 54.800 | 48.312 | 57.241 | 13.603 | 5.994 | 4.299 | 0.898 | 1.626 | 1.213 | 0.849 | 0.933 | 135.303 |
| 2007 | 10.817 | 94.250 | 83.631 | 163.294 | 61.207 | 87.541 | 92.126 | 23.238 | 11.728 | 7.319 | 2.593 | 4.961 | 2.302 | 1.420 | 158.917 |
| 2008 | 10.427 | 38.830 | 90.932 | 79.745 | 107.644 | 59.656 | 62.194 | 54.345 | 18.130 | 8.240 | 5.157 | 2.680 | 2.630 | 1.178 | 151.780 |
| 2009 | 5.431 | 21.856 | 35.221 | 31.914 | 18.826 | 22.725 | 10.425 | 9.213 | 9.549 | 2.238 | 1.033 | 0.768 | 0.406 | 0.298 | 46.332 |
| 2010 | 1.476 | 8.843 | 22.674 | 29.492 | 24.293 | 14.419 | 17.407 | 10.045 | 7.576 | 8.896 | 1.764 | 1.105 | 0.672 | 0.555 | 43.533 |
| 2011 | 0.521 | 9.357 | 24.621 | 20.046 | 22.869 | 23.706 | 13.749 | 16.967 | 10.039 | 7.623 | 7.745 | 1.441 | 0.618 | 0.785 | 49.446 |
| $2012^{*}$ | 0.403 | 17.827 | 89.432 | 51.257 | 43.079 | 51.224 | 41.846 | 34.653 | 27.215 | 24.946 | 15.473 | 13.575 | 2.595 | 0.253 | 125.369 |
| 2013 | 6.888 | 46.848 | 24.833 | 35.070 | 17.250 | 18.550 | 19.032 | 21.821 | 15.952 | 15.804 | 10.081 | 9.775 | 6.722 | 2.486 | 72.058 |
| 2014 | 0.000 | 3.537 | 53.241 | 50.609 | 70.044 | 34.393 | 22.084 | 22.138 | 13.298 | 17.761 | 7.974 | 4.461 | 2.862 | 1.746 | 94.975 |
| 2015 | 0.089 | 6.024 | 29.89 | 53.573 | 43.501 | 43.015 | 15.533 | 10.76 | 8.664 | 8.161 | 6.981 | 2.726 | 2.467 | 1.587 | 69.729 |

* Includes both the catches and the herring that died in the mass mortality in the winter 2012/13 in Kolgrafafjörður

Table 11.2.2.2. Icelandic summer-spawning herring. The mean weight ( g ) at age from the commercial catch ( 1981 refers to season 1981/1982 etc.).

| Year \age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 110 | 179 | 241 | 291 | 319 | 339 | 365 | 364 | 407 | 389 | 430 | 416 | 416 | 416 |
| $1976$ | $103$ | $189$ | 243 | 281 | 305 | 335 | 351 | 355 | 395 | 363 | 396 | 396 | 396 | 396 |
| 1977 | 84 | 157 | 217 | 261 | 285 | 313 | 326 | 347 | 364 | 362 | 358 | 355 | 400 | 420 |
| $1978$ | 73 | 128 | 196 | 247 | 295 | 314 | 339 | 359 | 360 | 376 | 380 | 425 | 425 | 425 |
| 1979 | 75 | 145 | 182 | 231 | 285 | 316 | 334 | 350 | 367 | 368 | 371 | 350 | 350 | 450 |
| $1980$ | 69 | $115$ | 202 | 232 | 269 | 317 | 352 | 360 | 380 | 383 | 393 | 390 | 390 | 390 |
| 1981 | 61 | 141 | 190 | 246 | 269 | 298 | 330 | 356 | 368 | 405 | 382 | 400 | 400 | 400 |
| $1982$ | 65 | $141$ | $186$ | 217 | 274 | 293 | 323 | 354 | 385 | 389 | 400 | 394 | 390 | 420 |
| $1983$ | 59 | 132 | 180 | 218 | 260 | 309 | 329 | 356 | 370 | 407 | 437 | 459 | 430 | 472 |
| $1984$ | 49 | 131 | 189 | 217 | 245 | 277 | 315 | 322 | 351 | 334 | 362 | 446 | 417 | 392 |
| $1985$ | 53 | 146 | 219 | 266 | 285 | 315 | 335 | 365 | 388 | 400 | 453 | 469 | 433 | 447 |
| 1986 | 60 | 140 | 200 | 252 | 282 | 298 | 320 | 334 | 373 | 380 | 394 | 408 | 405 | 439 |
| $1987$ | 60 | 168 | 200 | 240 | 278 | 304 | 325 | 339 | 356 | 378 | 400 | 404 | 424 | 430 |
| 1988 | 75 | 157 | 221 | 239 | 271 | 298 | 319 | 334 | 354 | 352 | 371 | 390 | 408 | 437 |
| $1989$ | 63 | 130 | 206 | 246 | 261 | 290 | 331 | 338 | 352 | 369 | 389 | 380 | 434 | 409 |
| 1990 | 80 | 127 | 197 | 245 | 272 | 285 | 305 | 324 | 336 | 362 | 370 | 382 | 375 | 378 |
| 1991 | 74 | 135 | 188 | 232 | 267 | 289 | 304 | 323 | 340 | 352 | 369 | 402 | 406 | 388 |
| $1992$ | 68 | 148 | 190 | 235 | 273 | 312 | 329 | 339 | 355 | 382 | 405 | 377 | 398 | 398 |
| 1993 | 66 | 145 | 211 | 246 | 292 | 324 | 350 | 362 | 376 | 386 | 419 | 389 | 389 | 389 |
| 1994 | 66 | 134 | 201 | 247 | 272 | 303 | 333 | 366 | 378 | 389 | 390 | 412 | 418 | 383 |
| 1995 | 68 | 130 | 183 | 240 | 277 | 298 | 325 | 358 | 378 | 397 | 409 | 431 | 430 | 467 |
| 1996 | 75 | 139 | 168 | 212 | 258 | 289 | 308 | 325 | 353 | 353 | 377 | 404 | 395 | 410 |
| 1997 | 63 | 131 | 191 | 233 | 269 | 300 | 324 | 341 | 355 | 362 | 367 | 393 | 398 | 411 |
| 1998 | 52 | 134 | 185 | 238 | 264 | 288 | 324 | 340 | 348 | 375 | 406 | 391 | 426 | 456 |
| $1999$ | 74 | 137 | 204 | 233 | 268 | 294 | 311 | 339 | 353 | 362 | 378 | 385 | 411 | 422 |
| 2000 | 62 | 159 | 217 | 268 | 289 | 325 | 342 | 363 | 378 | 393 | 407 | 425 | 436 | 430 |
| 2001 | 74 | 139 | 214 | 244 | 286 | 296 | 324 | 347 | 354 | 385 | 403 | 421 | 421 | 433 |


| Year \age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 85 | 161 | 211 | 258 | 280 | 319 | 332 | 354 | 405 | 396 | 416 | 433 | 463 | 460 |
| 2003 | 72 | 156 | 189 | 229 | 260 | 283 | 309 | 336 | 336 | 369 | 394 | 378 | 412 | 423 |
| 2004 | 84 | 149 | 213 | 248 | 280 | 315 | 331 | 349 | 355 | 379 | 388 | 412 | 419 | 425 |
| 2005 | 106 | 170 | 224 | 262 | 275 | 298 | 324 | 335 | 335 | 356 | 372 | 394 | 405 | 413 |
| 2006 | 107 | 189 | 234 | 263 | 290 | 304 | 339 | 349 | 369 | 416 | 402 | 413 | 413 | 467 |
| 2007 | 93 | 158 | 221 | 245 | 261 | 277 | 287 | 311 | 339 | 334 | 346 | 356 | 384 | 390 |
| 2008 | 105 | 174 | 232 | 275 | 292 | 307 | 315 | 327 | 345 | 366 | 377 | 372 | 403 | 434 |
| 2009 | 113 | 190 | 237 | 274 | 304 | 318 | 326 | 335 | 342 | 360 | 372 | 394 | 409 | 421 |
| 2010 | 87 | 204 | 243 | 271 | 297 | 315 | 329 | 335 | 341 | 351 | 367 | 366 | 405 | 416 |
| 2011 | 97 | 187 | 245 | 283 | 309 | 328 | 343 | 352 | 356 | 364 | 375 | 386 | 378 | 432 |
| 2012 | 65 | 206 | 244 | 282 | 301 | 320 | 333 | 344 | 350 | 359 | 364 | 367 | 373 | 391 |
| 2013 | 95 | 182 | 238 | 271 | 300 | 322 | 337 | 349 | 360 | 365 | 362 | 375 | 377 | 394 |
| 2014 |  | 202 | 259 | 288 | 306 | 328 | 346 | 354 | 362 | 366 | 367 | 380 | 383 | 403 |
| 2015 | 107 | 203 | 249 | 275 | 299 | 313 | 329 | 347 | 352 | 358 | 361 | 368 | 380 | 378 |

Table 11.2.2.3. Icelandic summer-spawning herring. Proportion mature at age ( 1981 refers to season 1981/1982 etc.).

| YEAR $\backslash$ AGE | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5 +}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 0 | 0.27 | 0.97 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1976 | 0 | 0.13 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1977 | 0 | 0.02 | 0.87 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1978 | 0 | 0.04 | 0.78 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1979 | 0 | 0.07 | 0.65 | 0.98 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1980 | 0 | 0.05 | 0.92 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1981 | 0 | 0.03 | 0.65 | 0.99 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1982 | 0.02 | 0.05 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1983 | 0 | 0 | 0.64 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1984 | 0 | 0.01 | 0.82 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1985 | 0 | 0 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| $1986-2015$ | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 11.3.2.1. Icelandic summer-spawning herring. Natural mortality-at-age where the deviation from the fixed $\mathbf{M}=0.1$ is due to the Ichthyophonus infection ( 1981 refers to season 1981/1982 etc.).

| YEAR $\backslash$ AGE | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | 11 | 12 | 13 | 14 | $\mathbf{1 5 +}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $1987-2008$ | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 |
| 2009 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 |
| 2010 | 0.458 | 0.74 | 0.74 | 0.69 | 0.63 | 0.6 | 0.58 | 0.57 | 0.56 | 0.54 | 0.53 | 0.52 | 0.56 | 0.58 |
| $2011-2015$ | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 |

Table 11.3.2.2. Model settings and results of model parameters from the final NFT-Adapt run in 2016 for Icelandic summer spawning herring.

| VPA VERSION 3.3.0 | InPUT FILE: D: $\backslash$ NFT $\backslash$ VPA $\backslash 2016 \backslash$ RUN1 $\backslash$ RUN1.DAT |
| :--- | :--- | :--- |
| Model ID: Run0 - final last year <br> - test. | Date of Run: 11-APR-2016 Time of Run: 14:58 |

Levenburg-Marquardt Algorithm Completed 5 Iterations
Residual Sum of Squares $=51.5991$

| Number of Residuals $=224$ | Number of Years $=29$ |
| :--- | :--- |
| Number of Parameters $=9$ | Number of Ages $=11$ |
| Degrees of Freedom $=215$ | First Year $=1987$ |
| Mean Squared Residual $=0.239996$ | Youngest Age $=3$ |
| Standard Deviation $=0.489894$ | Oldest True Age $=12$ |

Number of Survey Indices Available $=10$
Number of Survey Indices Used in Estimate $=8$
VPA Classic Method - Auto Estimated Q's

| Stock Numbers Predicted in Terminal Year Plus One (2016): |  |  |  |
| :---: | :---: | :---: | :---: |
| Age | Stock Predicted | Std. Error | CV |
| 4 | 165038.2370 .82410 | $8 \mathrm{E}+05 \quad 0.4993$ | $4 \mathrm{E}+00$ |
| 5 | $98590.480 \quad 0.3890$ | 1E+05 0.3946 | 3E+00 |
| 6 | 207453.606 | $0.713928 \mathrm{E}+05$ | $0.344139 \mathrm{E}+00$ |
| 7 | 168231.145 | $0.535604 \mathrm{E}+05$ | $0.318374 \mathrm{E}+00$ |
| 8 | 169986.144 | $0.528620 \mathrm{E}+05$ | $0.310979 \mathrm{E}+00$ |
| 9 | 64530.831 | $0.200462 \mathrm{E}+05$ | $0.310645 \mathrm{E}+00$ |
| 10 | 38048.067 | $0.122208 \mathrm{E}+05$ | $0.321194 \mathrm{E}+00$ |
| 11 | 31940.033 | $0.105191 \mathrm{E}+05$ | $0.329338 \mathrm{E}+00$ |
| 12 | 10337.521 | $0.653617 \mathrm{E}+04$ | $0.632277 \mathrm{E}+00$ |
| Catchability values for each survey used in estimate: |  |  |  |
| INDEX | EX Catchability | Std. Error | CV |
| 1 | $0.995048 \mathrm{E}+00$ | $0.100122 \mathrm{E}+00$ | $0.100620 \mathrm{E}+00$ |
| 2 | $0.126132 \mathrm{E}+01$ | $0.119572 \mathrm{E}+00$ | $0.947991 \mathrm{E}-01$ |
| 3 | $0.125871 \mathrm{E}+01$ | $0.872105 \mathrm{E}-01$ | $0.692855 \mathrm{E}-01$ |
| 4 | $0.133669 \mathrm{E}+01$ | $0.913517 \mathrm{E}-01$ | $0.683419 \mathrm{E}-01$ |
| 5 | $0.145707 \mathrm{E}+01$ | $0.114556 \mathrm{E}+00$ | $0.786210 \mathrm{E}-01$ |
| 6 | $0.165918 \mathrm{E}+01$ | $0.152462 \mathrm{E}+00$ | $0.918903 \mathrm{E}-01$ |
| 7 | $0.175855 \mathrm{E}+01$ | $0.202623 \mathrm{E}+00$ | $0.115221 \mathrm{E}+00$ |
| 8 | $0.169599 \mathrm{E}+01$ | $0.198723 \mathrm{E}+00$ | $0.117172 \mathrm{E}+00$ |

-- Non-Linear Least Squares Fit --

Maximum Marquadt Iterations $=100$
Scaled Gradient Tolerance $=6.055454 \mathrm{E}-05$
Scaled Step Tolerance $\quad=1.000000 \mathrm{E}-18$
Relative Function Tolerance $=1.000000 \mathrm{E}-18$
Absolute Function Tolerance $=4.930381 \mathrm{E}-32$
Reported Machine Precision $=2.220446 \mathrm{E}-16$
VPA Method Options:

- Catchability Values Estimated as an Analytic Function of N
- Catch Equation Used in Cohort Solution
- Plus Group Forward Calculation Method Used
- Arithmetic Average Used in F-Oldest Calculation
- F-Oldest Calculation in Years Prior to Terminal Year

Uses Fishing Mortality in Ages 8 to 11

- Calculation of Population of Age 3 In Year 2016
$=$ Geometric Mean of First Age Populations Year Range Applied = 1991 to 2012
- Survey Weight Factors Were Used

Stock Estimates: Age 4, Age 5, Age 6, Age 7, Age 8, Age 9, Age 10, Age 11, Age 12

Full F in Terminal Year $\quad=0.2692$
F in Oldest True Age in Terminal Year $=0.3076$

Full F Calculated Using Classic Method

F in Oldest True Age in Terminal Year has been
Calculated in Same Manner as in All Other Years

| Age | Input Partial | Calc Partial Fishing Used In |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Recruitment | Mortality | Full F |  | Comments |
| 3 | 0.500 | 0.061 | 0.0341 | NO | Stock Estimate in $\mathrm{T}+1$ |
| 4 | 0.800 | 0.453 | 0.2528 | NO | Stock Estimate in $\mathrm{T}+1$ |
| 5 | 1.000 | 0.393 | 0.2192 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 6 | 1.000 | 0.393 | 0.2195 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 7 | 1.000 | 0.386 | 0.2153 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 8 | 1.000 | 0.369 | 0.2058 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 9 | 1.000 | 0.426 | 0.2377 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 10 | 1.000 | 0.411 | 0.2291 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 11 | 1.000 | 1.000 | 0.5580 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 12 | 1.000 | 0.551 | 0.3076 |  | F-Oldest |

Table 11.3.2.3. Icelandic summer spawners stock estimates (from NFT-Adapt in 2016) in numbers (millions) by age (years) at January $1^{\text {st }}$ during 1987-2016.


|  |  |  | 449.7 | 313.7 | 276.5 | 205.5 | 158.8 | 108.5 | 113.3 | 23.4 | 32.5 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | 613.47 | 462.71 | 2 | 1 | 6 | 4 | 7 | 0 | 6 | 6 | 4 | 2758 |
|  |  |  | 205.5 | 205.2 | 149.7 | 141.3 | 102.3 |  |  | 59.3 | 29.6 |  |
| 2011 | 605.22 | 286.73 | 2 | 0 | 4 | 0 | 1 | 82.42 | 56.35 | 8 | 9 | 1924 |
|  |  |  | 236.0 | 166.9 | 163.9 | 112.9 | 114.7 |  |  | 43.7 | 70.5 |  |
| 2012 | 424.97 | 538.73 | 5 | 2 | 5 | 9 | 9 | 76.47 | 65.04 | 5 | 4 | 2014 |
|  |  |  | 402.5 | 164.9 | 110.1 |  |  |  |  | 35.2 | 73.1 |  |
| 2013 | 459.60 | 367.58 | 7 | 6 | 8 | 99.80 | 62.61 | 71.02 | 43.41 | 4 | 8 | 1890 |
|  |  |  | 309.0 | 330.9 | 132.8 |  |  |  |  | 24.3 | 70.5 |  |
| 2014 | 158.78 | 371.36 | 0 | 4 | 7 | 82.09 | 72.24 | 35.98 | 49.13 | 1 | 4 | 1637 |
|  |  |  | 285.4 | 231.5 | 232.9 |  |  |  |  | 27.6 | 69.6 |  |
| 2015 | 188.72 | 140.30 | 7 | 5 | 8 | 87.61 | 53.33 | 44.39 | 19.96 | 3 | 6 | 1382 |
|  |  |  |  | 207.4 | 168.2 | 169.9 |  |  |  | 10.3 | 74.9 |  |
| $2016^{*}$ | 464.22 | 165.04 | 98.59 | 5 | 3 | 9 | 64.53 | 38.05 | 31.94 | 4 | 7 | 1678 |

* Number-at-age 3 in 2016 is predicted from an survey index of number-at-age 1 in 2013 (see section 11.6.1)

Table 11.3.2.4. Estimated fishing mortality-at-age of Icelandic summer-spawning herring (from NFT-Adapt in 2016) by age (years) during 1987-2015 (referring to the autumn of the fishing season) and weighed average $F$ by numbers for age 5-10.

| Year \AGE | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ | WF5-10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 0.006 | 0.049 | 0.236 | 0.295 | 0.356 | 0.598 | 0.468 | 0.485 | 0.516 | 0.517 | 0.517 | 0.347 |
| 1988 | 0.019 | 0.096 | 0.131 | 0.412 | 0.547 | 0.654 | 0.988 | 0.764 | 0.704 | 0.777 | 0.506 | 0.266 |
| 1989 | 0.055 | 0.124 | 0.234 | 0.345 | 0.432 | 0.335 | 0.356 | 0.550 | 0.674 | 0.479 | 0.111 | 0.322 |
| 1990 | 0.053 | 0.170 | 0.216 | 0.353 | 0.477 | 0.542 | 0.586 | 0.431 | 0.471 | 0.508 | 0.071 | 0.399 |
| 1991 | 0.117 | 0.222 | 0.369 | 0.301 | 0.392 | 0.640 | 0.309 | 0.592 | 0.466 | 0.502 | 0.055 | 0.435 |
| 1992 | 0.101 | 0.210 | 0.242 | 0.408 | 0.372 | 0.460 | 0.649 | 0.613 | 0.464 | 0.547 | 0.023 | 0.414 |
| 1993 | 0.088 | 0.249 | 0.218 | 0.179 | 0.316 | 0.272 | 0.400 | 0.344 | 0.421 | 0.359 | 0.011 | 0.247 |
| 1994 | 0.227 | 0.254 | 0.289 | 0.291 | 0.301 | 0.346 | 0.547 | 0.569 | 0.572 | 0.509 | 0.090 | 0.311 |
| 1995 | 0.114 | 0.339 | 0.282 | 0.332 | 0.398 | 0.403 | 0.419 | 0.548 | 0.731 | 0.525 | 0.154 | 0.341 |
| 1996 | 0.096 | 0.295 | 0.328 | 0.335 | 0.411 | 0.328 | 0.476 | 0.383 | 0.799 | 0.496 | 0.348 | 0.358 |
| 1997 | 0.054 | 0.236 | 0.285 | 0.202 | 0.243 | 0.231 | 0.276 | 0.494 | 0.232 | 0.308 | 1.033 | 0.247 |
| 1998 | 0.158 | 0.322 | 0.250 | 0.127 | 0.322 | 0.266 | 0.351 | 0.262 | 0.194 | 0.268 | 0.571 | 0.275 |
| 1999 | 0.110 | 0.259 | 0.411 | 0.252 | 0.267 | 0.390 | 0.298 | 0.385 | 0.336 | 0.352 | 0.712 | 0.370 |
| 2000 | 0.159 | 0.373 | 0.348 | 0.365 | 0.258 | 0.233 | 0.321 | 0.212 | 0.311 | 0.269 | 0.665 | 0.326 |
| 2001 | 0.093 | 0.337 | 0.420 | 0.333 | 0.424 | 0.345 | 0.366 | 0.375 | 0.303 | 0.347 | 0.433 | 0.397 |
| 2002 | 0.055 | 0.199 | 0.267 | 0.406 | 0.365 | 0.601 | 0.567 | 0.305 | 0.311 | 0.446 | 0.865 | 0.388 |
| 2003 | 0.205 | 0.249 | 0.223 | 0.232 | 0.304 | 0.394 | 0.270 | 0.212 | 0.337 | 0.303 | 0.227 | 0.253 |
| 2004 | 0.085 | 0.182 | 0.222 | 0.213 | 0.153 | 0.155 | 0.245 | 0.248 | 0.455 | 0.276 | 0.250 | 0.213 |
| 2005 | 0.024 | 0.067 | 0.206 | 0.224 | 0.197 | 0.175 | 0.201 | 0.321 | 0.257 | 0.239 | 0.188 | 0.213 |
| 2006 | 0.138 | 0.247 | 0.105 | 0.110 | 0.127 | 0.127 | 0.111 | 0.122 | 0.076 | 0.109 | 0.136 | 0.115 |
| 2007 | 0.119 | 0.121 | 0.258 | 0.154 | 0.270 | 0.280 | 0.296 | 0.294 | 0.281 | 0.288 | 0.324 | 0.244 |
| 2008 | 0.046 | 0.145 | 0.148 | 0.247 | 0.189 | 0.272 | 0.232 | 0.349 | 0.307 | 0.290 | 0.268 | 0.213 |
| 2009 | 0.036 | 0.059 | 0.076 | 0.051 | 0.082 | 0.050 | 0.064 | 0.063 | 0.071 | 0.062 | 0.055 | 0.065 |
| 2010 | 0.021 | 0.072 | 0.095 | 0.110 | 0.072 | 0.118 | 0.086 | 0.095 | 0.107 | 0.101 | 0.097 | 0.096 |
| 2011 | 0.016 | 0.095 | 0.108 | 0.124 | 0.182 | 0.108 | 0.191 | 0.137 | 0.153 | 0.147 | 0.106 | 0.136 |
| 2012 | 0.045 | 0.191 | 0.258 | 0.315 | 0.396 | 0.490 | 0.380 | 0.466 | 0.513 | 0.462 | 0.280 | 0.360 |
| 2013 | 0.113 | 0.074 | 0.096 | 0.116 | 0.194 | 0.223 | 0.454 | 0.269 | 0.480 | 0.356 | 0.317 | 0.164 |
| 2014 | 0.024 | 0.163 | 0.189 | 0.251 | 0.317 | 0.331 | 0.387 | 0.489 | 0.476 | 0.421 | 0.145 | 0.266 |
| 2015 | 0.034 | 0.253 | 0.219 | 0.220 | 0.215 | 0.206 | 0.238 | 0.229 | 0.558 | 0.308 | 0.108 | 0.219 |

Table 11.3.2.5. Summary table from NFT-Adapt run in 2016 for Icelandic summer spawning herring.

| Year | RECRUITS, AGE 3 (MILLIONS) | BIomASS AGE $3+$ (KT) | $\begin{aligned} & \text { SSB } \\ & \text { (KT) } \end{aligned}$ | LANDINGS AGE $3+$ (KT) | $\begin{gathered} \text { Yield/S } \\ \text { SB } \end{gathered}$ | WFAGE 510 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 530 | 504 | 384 | 75 | 0.20 | 0.35 |
| 1988 | 271 | 495 | 423 | 93 | 0.22 | 0.27 |
| 1989 | 448 | 459 | 386 | 101 | 0.26 | 0.32 |
| 1990 | 301 | 410 | 350 | 104 | 0.30 | 0.40 |
| 1991 | 843 | 424 | 310 | 107 | 0.34 | 0.44 |
| 1992 | 1036 | 503 | 344 | 107 | 0.31 | 0.41 |
| 1993 | 638 | 547 | 425 | 103 | 0.24 | 0.25 |
| 1994 | 695 | 555 | 442 | 134 | 0.30 | 0.31 |
| 1995 | 205 | 464 | 408 | 125 | 0.31 | 0.34 |
| 1996 | 183 | 350 | 310 | 96 | 0.31 | 0.36 |
| 1997 | 779 | 372 | 272 | 65 | 0.24 | 0.25 |
| 1998 | 326 | 371 | 302 | 86 | 0.28 | 0.28 |
| 1999 | 566 | 379 | 295 | 93 | 0.31 | 0.37 |
| 2000 | 410 | 397 | 314 | 100 | 0.32 | 0.33 |
| 2001 | 499 | 363 | 283 | 94 | 0.33 | 0.40 |
| 2002 | 1591 | 552 | 317 | 96 | 0.30 | 0.39 |
| 2003 | 1195 | 637 | 427 | 129 | 0.30 | 0.25 |
| 2004 | 818 | 706 | 552 | 112 | 0.20 | 0.21 |
| 2005 | 1180 | 838 | 623 | 102 | 0.16 | 0.21 |
| 2006 | 963 | 971 | 751 | 130 | 0.17 | 0.11 |
| 2007 | 878 | 906 | 733 | 158 | 0.22 | 0.24 |
| 2008 | 901 | 980 | 789 | 151 | 0.19 | 0.21 |
| 2009 | 783 | 1001 | 669 | 46 | 0.07 | 0.07 |
| 2010 | 613 | 759 | 467 | 43 | 0.09 | 0.10 |
| 2011 | 605 | 523 | 401 | 49 | 0.12 | 0.14 |
| 2012 |  |  |  |  |  |  |
| * | 425 | 557 | 445 | 125 | 0.28 | 0.36 |
| 2013 | 460 | 503 | 402 | 71 | 0.18 | 0.16 |
| 2014 | 159 | 483 | 421 | 95 | 0.23 | 0.27 |
| 2015 | 189 | 400 | 346 | 70 | 0.20 | 0.22 |
| 2016 | 464 | 413 | 318 |  |  |  |
| Mea n | 632 | 561 | 430 | 99 | 0.24 | 0.28 |

* The mass mortality of 52 thousand tons in Kolgrafafjörour in the winter 2012/13 is included in the landings, yield/SSB, and WF.
§ Number-at-age 3 in 2016 is predicted from an survey index of number-at-age 1 in 2013 (see section 11.6.1)

Table 11.3.2.6. The residuals from survey observations and NFT-Adapt 2016 results for Icelandic summer spawning herring (no surveys in 1987 and 1995) on 1st January.

| Year \AGE | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 |  |  |  |  |  |  |  |  |
| $1988$ | -0.167 | -0.226 | 0.130 | -0.288 | -0.675 | -0.231 | -0.146 | -0.421 |
| 1989 | -0.174 | -0.753 | -0.804 | 0.091 | 0.066 | -0.003 | 0.000 | 0.000 |
| 1990 | 0.540 | -0.303 | -0.237 | 0.022 | 0.489 | -0.368 | -0.001 | -0.002 |
| 1991 | -0.665 | -0.357 | -0.628 | -0.222 | 0.371 | 0.184 | 0.008 | -0.003 |
| 1992 | 0.442 | $0.406$ | 0.327 | -0.337 | -0.139 | 0.287 | -0.782 | 0.002 |
| 1993 | $-0.014$ | 0.151 | -0.053 | -0.120 | -0.456 | -0.071 | 0.000 | 0.109 |
| 1994 | -0.041 | 0.158 | 0.086 | -0.697 | -0.598 | 0.459 | -0.308 | -0.502 |
| $1995$ |  |  |  |  |  |  |  |  |
| 1996 | -0.206 | 0.625 | -0.138 | 0.090 | -0.203 | 0.373 | -0.005 | -0.146 |
| 1997 | 0.592 | -0.049 | 0.570 | 0.208 | 0.348 | 0.302 | 0.836 | 0.648 |
| 1998 | -0.099 | -0.515 | -0.508 | 0.322 | -0.084 | 0.078 | -0.102 | 0.502 |
| 1999 | 0.023 | 0.672 | 0.080 | -0.443 | -0.095 | -0.642 | -0.223 | -0.376 |
| 2000 | 0.609 | 0.077 | 0.606 | 0.213 | -0.340 | 0.468 | -0.059 | 0.475 |
| 2001 | 1.123 | 1.295 | 0.302 | 0.782 | -0.458 | -1.151 | -0.644 | -1.548 |
| $2002$ | $-0.356$ | -0.169 | 0.197 | 0.505 | 0.887 | 0.454 | 0.547 | -0.121 |
| 2003 | 0.347 | 0.362 | 0.144 | 0.650 | 0.833 | 1.233 | 1.525 | 0.806 |
| 2004 | 0.494 | 0.529 | 0.173 | -0.217 | 0.011 | -0.178 | -0.257 | -0.008 |
| 2005 | 0.053 | 0.199 | 0.177 | -0.232 | -0.609 | -0.686 | -1.154 | -0.516 |
| 2006 | -0.849 | -0.744 | 0.285 | 0.595 | 0.475 | 0.208 | 0.630 | 1.205 |
| 2007 | -0.217 | 0.131 | -0.364 | -0.222 | 0.170 | -0.504 | 0.374 | -0.089 |
| 2008 | -0.430 | -0.976 | -0.180 | -0.452 | 0.010 | 0.433 | 0.671 | 1.486 |
| 2009 | -1.374 | -0.514 | -0.730 | -0.062 | -0.397 | -0.327 | -0.718 | -0.846 |
| 2010 | -0.377 | -0.145 | 0.340 | -0.309 | 0.089 | -0.569 | -0.822 | -0.203 |
| 2011 | -0.017 | -0.125 | 0.166 | 0.392 | -0.395 | 0.574 | -0.866 | 0.391 |
| 2012 | 0.601 | 0.379 | 0.401 | 0.217 | 0.374 | -0.209 | 0.264 | -0.266 |
| 2013 | 0.759 | 0.218 | -0.220 | -0.148 | -0.024 | 0.058 | -0.253 | 0.009 |
| 2014 | -0.160 | -0.578 | -0.189 | -0.158 | 0.104 | 0.007 | 0.626 | 0.072 |
| 2015 | -0.436 | -0.087 | -0.111 | -0.184 | 0.387 | 0.235 | 0.160 | 0.112 |
| 2016 | 0.000 | 0.339 | 0.179 | 0.004 | -0.141 | -0.133 | -0.064 | 0.330 |

Table 11.6.1.1. The input data used for prognosis of the Icelandic summer-spawning herring in the 2016 assessment: the predicted weights, the selection pattern, $M$, proportion of $M$ before spawning, and the number-at-age derived from NFT-Adapt run.

| Age (year CLASS) | Mean WEIGHTS (KG) | M | MATURITY OGIVE | Selection <br> PATTERN | MORTALITY PROP. before spawning |  | Number-atAGE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | F | M | JAN. 1 ST 2016 |
| 3 (2013) | 0.173 | 0.10 | 0.200 | 0.201 | 0.000 | 0.500 | 464.2 |
| 4 (2012) | 0.248 | 0.10 | 0.850 | 0.782 | 0.000 | 0.500 | 165.0 |
| 5 (2011) | 0.284 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 98.6 |
| 6 (2010) | 0.304 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 207.5 |
| 7 (2009) | 0.323 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 168.2 |
| 8 (2008) | 0.334 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 170.0 |
| 9 (2007) | 0.346 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 64.5 |
| 10 (2006) | 0.360 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 38.0 |
| 11 (2005) | 0.364 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 31.9 |
| 12 (2004) | 0.368 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 10.3 |
| 13+ (2003+) | 0.371 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 75.0 |

Table 11.6.1.2. Icelandic summer-spawning herring. Short-term prediction for the 2016/2017 season where the basis is: SSB (2016): 318 kt ; Biomass age 3+ (2016): 393 kt (at spawning time); Catch(2015/16): 70 kt ; WF5-10(2015)=0.219. The fishery has been managed on basis of $F 0.1=0.22$ for over 20 years. SSB is in the spawning seasons, which is approximately the beginning of the subsequent fishing season. Catches and SSB are in thousands tons.

| Rationale | LANDINGS | BASIS | F | SSB | \%SSB Change <br> 1) | \% TAC CHANGE <br> 2) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2016/2017 |  | 2016/2017 | 2017 |  |  |
| MSY <br> approach | 63 | Fmsy | 0.22 | 303 | -7 | -12 |
| F0.1 | 63 | $\mathrm{F}_{0.1}=\mathrm{F}_{\mathrm{pa}}=0.22$ | 0.22 | 303 | -7 | -12 |
| Zero catch | 0 | $\mathrm{F}=0$ | 0.00 | 358 | 9 |  |
| Status quo | 63 | F(2013) | 0.16 | 303 | -7 | -12 |
| Fmult | 6 | $0.1 \times$ (F0.1) | 0.02 | 353 | 8 | -1011 |
|  | 15 | $0.25 \times(\mathrm{F} 0.1)$ | 0.05 | 345 | 6 | -355 |
|  | 33 | $0.5 \times(\mathrm{F} 0.1)$ | 0.11 | 329 | 1 | -113 |
|  | 49 | $0.75 \times(\mathrm{F} 0.1)$ | 0.17 | 315 | -3 | -42 |
|  | 57 | $0.9 \times(\mathrm{F} 0.1)$ | 0.20 | 308 | -6 | -22 |
|  | 68 | $1.1 \times(\mathrm{F} 0.1)$ | 0.24 | 299 | -9 | -4 |
|  | 75 | $1.25 \times(\mathrm{F} 0.1)$ | 0.27 | 292 | -11 | 7 |
|  | 89 | $1.5 \times(\mathrm{F} 0.1)$ | 0.33 | 280 | -16 | 22 |

1) SSB 2017 relative to SSB 2016.
2) TAC $2016 / 17$ relative to landings $2015 / 16$.

Table 11.6.1.3. Icelandic summer-spawning herring. Medium term prediction where the basis is: SSB(2016): 318 kt ; Biomass age 3+ (2016): 393 kt (at spawning time); Catch(2015/16): 70 kt ; WF5$10(2015)=0.219$. The prognosis of the Icelandic summer spawning herring for the next fishing season (2016/2017; grey shaded) and the two subsequent seasons under five different options ( $F_{0.1}=0.22$, constant TAC of $60 \mathrm{kt}, 70 \mathrm{kt}, 80 \mathrm{kt}$ and 90 kt ) from the final NFT-Adapt run in 2016. SSBs are in the spawning seasons, which is approximately the beginning of the subsequent fishing season.

| $2016 / 2017$ | SPAWNING 2017 | $2017 / 2018$ | SPAWNING 2018 | $2018 / 2019$ | SPAWNING 2019 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TAC | F | BIOMASS 3+ | SSB (KT) | TAC | F | BIOMASS 3+ | SSB (KT) | TAC | F | BIOMASS 3+ | SSB (KT) |
| $(\mathrm{kt})$ | $(5-10)$ | $(\mathrm{kt})$ |  | $(\mathrm{kt})$ | $(5-10)$ | $(\mathrm{kt})$ | $(\mathrm{kt})$ | $(\mathrm{kt})$ | $(5-10)$ | $(\mathrm{kt})$ | $(\mathrm{kt})$ |
| 63 | 0.22 | 386 | 303 | 60 | 0.22 | 431 | 321 | 63 | 0.22 | 475 | 358 |
| 60 | 0.21 | 389 | 306 | 60 | 0.22 | 433 | 322 | 60 | 0.21 | 536 | 414 |
| 70 | 0.25 | 379 | 297 | 70 | 0.27 | 413 | 304 | 70 | 0.26 | 452 | 335 |
| 80 | 0.29 | 370 | 288 | 80 | 0.33 | 394 | 286 | 80 | 0.33 | 424 | 309 |
| 90 | 0.33 | 361 | 279 | 90 | 0.39 | 375 | 268 | 90 | 0.41 | 395 | 282 |

### 11.20 Figures



Figure 11.1.2.1. The survey tracks of five acoustic surveys on Icelandic summer-spawning herring in Sept.-Oct. 2016 (Surveys D4-2015 and D5-2015 on juveniles; blue line), October 2015 (D7-2015 on juveniles and adults; orange), January 2016 (B1-2016 on adults; red line) and in March 2016 (A4-2016 on adults; light blue line) and locations of the areas that are referred to in the text.


Figure 11.1.2.2 Total biomass index for Icelandic summer-spawning herring from the acoustic surveys for ages $3+$ in the areas east and west of $18^{\circ} \mathrm{W}$ (except in 2011 and 2012 where fish outside Breiðafjörður was set to the eastern part), combined over all areas and age 3-10 which are used in tuning of the analytical assessment. The years in the plot (1973-2015) refer to the autumn of the fishing seasons.


Figure 11.1.3.1. The prevalence of Ichthyophonus infection for the different year classes of Icelandic summer-spawning herring in Breiðafjörður and west of Iceland as estimated in the autumns 20082015.


Figure 11.2.1. Icelandic summer spawning herring. Seasonal total landings (in thousand tonnes) during 1947-2015, referring to the autumns, by different fishing gears (from 1975-2015).


Figure 11.2.2. The distribution of the fishery (in tonnes) of Icelandic summer spawning herring during the fishing season 2015/16, including the bycatch in the mackerel fishery in June-September 2015, where the stars indicate the location of catch samples.


Figure 11.2.2.1. Proportion of the different age groups of Icelandic summer-spawning herring to the total catches (biomass) as observed in 2015/2016 fishing season (June 2015-Februay 2016), predicted in the 2015 assessment (ICES 2015) for the 2015/2016 fishing season, and the summer catches in June-September 2015 compared with the age composition in the stock according to the acoustic measurements in the winter 2015/2016.


Figure 11.3.1.1. Icelandic summer-spawning herring. Catch curves by year classes 1984-2011. Grey lines correspond to $\mathrm{Z}=0.4$. Note that the mass mortality in Kolgrafafjörour is added to the catches in 2012.


Figure 11.3.1.2. Icelandic summer spawning herring. Catch curves from survey data by year classes 1984-2011. Grey lines correspond to $\mathrm{Z}=0.4$.


Figure 11.3.2.1. Icelandic summer-spawning herring. The catchability ( $\pm 2$ SE) and its CV for the acoustic surveys used in the final Adapt run in 2016 (1987-2015) compare to the assessment in 2015.


Figure 11.3.2.2. Icelandic summer-spawning herring. Comparisons of the final NFT-Adapt runs in 2016 and in 2015 concerning (a) biomass of age 3-12, (b) biomass of age 4-12, (c) number-at-age 3, and N-weighed F for age 5-10. Note that the mass mortality in Kolgrafafjörður in the winter 2012/13 is included in weighed $F$ for that year (WF5-10 without the mass mortality was $\sim 0.22$ ) and the num-ber-at-age 3 in 2015 from Adapt 2015 was geometric mean while number-at-age 3 in 2016 was predicted from juvenile index (see section 11.6.1).


Figure 11.3.2.3. Icelandic summer spawning herring. Residuals of NFT-Adapt run in 2016 from survey observations (moved to 1st January). Filled bubbles are positive and open negative. Max bubble = 1.55


Figure 11.3.2.4. Icelandic summer spawning herring. Retrospective pattern from NFT-Adapt in 2016 in spawning-stock biomass (the top panel), $\mathbf{N}$ weighted $\mathrm{F} 5-10$ (middle panel) and recruitment as number-at-age 3 (lowest panel).


Figure 11.3.2.5. Icelandic summer-spawning herring. Observed vs.. predicted survey values from NFT-Adapt run in 2016 for ages $4-11$ with respect to numbers (upper) and biomass (lower).


Figure 11.3.2.6. Icelandic summer-spawning herring. Comparison of SSB (upper panel) and recruitment (lower panel) from the final NFT-Adapt run in 2016 (dark blue circles) and TASACS run in 2016 (light blue triangles).


Figure 11.3.2.7. Icelandic summer-spawning herring. Comparison of number-at-age on Jan. 1st. 2015 from the final NFT model runs in 2015 and 2016 assessments.


Figure 11.6.1.1. Icelandic summer spawning herring. The mean weight-at-age for age groups 3-12 (+ group) in 1987-2006, 2009-2013, in the catches in the winter 2015/2016, predicted weights for the winter 2015/2016 in the 2015 assessment (ICES 2015) and finally predicted weights for the autumn 2016 from the weights in 2015, which was used in the stock prognosis.


Figure 11.6.1.2. Icelandic summer spawning herring. The selection pattern for age groups 3-12 (+ group) for the years 2013-2015, the average selection across these three years, the selection used in 2015, and the selection used in the prognosis 2016 (three years average for age 3 and 4, but fixed at 1.0 for age 4+).


Figure 11.6.2.1. Icelandic summer spawning herring. The predicted biomass contribution of the different year classes to the catches in the fishing season 2016/2017 (total catch of 63 thousand tons).

## 12 Capelin in the Iceland-East Greenland-Jan Mayen area

### 12.1 Stock description and management units

See stock-annex.

### 12.2 Fishery-independent abundance surveys

The capelin stock in Iceland-East Greenland-Jan Mayen area has been assessed by acoustics annually since 1978. The surveys have taken place in autumn (SeptemberDecember) and in winter (January-February). An overview is given in the stock annex.

### 12.2.1 Surveys in autumn 2015

Two autumn surveys were conducted in 2015 with the aim of assessing both the immature and the maturing part of the stock. Since 2010 the autumn surveys have started in September, a month earlier than in previous years.

### 12.2.1.1 Autumn survey during 16. September - 4. October.

The survey area was on and along the shelf edge off East Greenland from about $73^{\circ}$ $30^{\prime} \mathrm{N}$ to about $65^{\circ} 30^{\prime} \mathrm{N}$ and between $16^{\circ}$ and $30^{\circ} \mathrm{W}$ including the Greenland Strait and the slope off western and north Iceland to about $16^{\circ} \mathrm{W}$ (Bardarson and Jonsson 2016a). Weather conditions during the survey were adverse but for the first few days and the survey had to be discontinued several times because of storms. Furthermore, drift ice in the northern part of the surveyed area (north of $72^{\circ} \mathrm{N}$ ) restricted the coverage in that region. Both the drift ice and storms delayed the progress of the cruise.
Figure 12.2 .1 shows the cruise tracks, distribution and relative density of the capelin of all ages and maturity stages. Immature capelin was found in unusually small numbers ( 6.2 billion, Tables 12.2.1 and 12.2.2 and Figure 12.2.2), mainly in the southwestern part of the surveyed area. Further north along the Greenland shelf up to $73^{\circ} \mathrm{N}$ older, maturing capelin predominated. No capelin was recorded off N -Iceland east of $21^{\circ} \mathrm{W}$. The distribution of the capelin was very westerly both for the 1-group and older capelin as it was in recent years (2010-2014) while unlike 2014 now no capelin was recorded in the more traditional areas north of Iceland. In this survey around 550 thousand t of mature capelin were estimated (Tables 12.2.1-12.2.3).

The estimates of both mature and immature capelin are considered to be minimum estimates (likely underestimates) because the survey did not reach the edge of their respective distributions. The edge of the mature capelin stock was not reached towards north and west, and the edge of the immature part of the stock was not reached towards west and south (Figure 12.2.2).

On the basis of the estimate of the maturing part of the stock the Marine Research Institute recommended an intermediate TAC of 44 thousand $t$ for the fishing season $2015 / 2016$. This recommendation was in accordance with existing HCR and management plan between Iceland, Norway and Greenland.

### 12.2.1.2 Autumn survey during 17-29. November

This survey was limited by adverse weather and fast moving drift ice (Bardarson 2016). The survey covered the Denmark Strait and the slope off western and north Iceland from about $29^{\circ} 30^{\prime} \mathrm{W}$ to $17^{\circ} \mathrm{W}$ and from $66^{\circ} 10^{\prime} \mathrm{N}$ (in western part) towards $68^{\circ} 30^{\prime} \mathrm{N}$ (eastern part) (Figure 12.2.1). At the beginning of the survey the edge of the drift ice was close to the centre of Denmark Strait but soon the ice started to drift southwards
into the Denmark Strait, hence it was attempted to measure that region as early as possible eastwards through the Denmark Strait. The 20th November a southward moving protrusion of the drift ice hindered further eastward progression of the survey, limiting the exploration of that region to the shelf edge north of Westfjords. Subsequently the shelf edge north of Iceland was surveyed to the east of Kolbeinsey Ridge and then northwards to about $68^{\circ} 30^{\prime} \mathrm{N}$ where further exploration had to stop due to a heavy storm. Finally, an attempt was made to cover the Denmark Strait again. This second attempt was restricted because of strong winds. The high densities that were found in the area during the first part of the survey were not found, suggesting that they were missed due to the wind conditions or because the fish had moved. The weather forecasts gave no hope for working conditions in the region for at least the next week, consequently the survey was ended and the ship returned to Reykjavik harbour the 29th November.

Capelin was only observed west of $23^{\circ} \mathrm{W}$. About 6.1 billion of immature capelin were measured mainly in the southwestern part of the surveyed area, while it was mixed with mature capelin further northeast. Mature capelin predominated along the shelf edge northwest of the Westfjords peninsula, just out of the Kögurgrunn and Hali shelf areas. In this survey about 295 thousand t of mature capelin was measured. No capelin was found on the transect furthest to the west in Denmark Strait nor along the continental shelf edge north of north-Iceland

Given the limitations of coverage due to weather and ice conditions this survey estimate was not used for TAC advice.

### 12.2.2 Surveys in winter 2016

Two winter surveys were conducted in year 2016 with the aim of assessing the maturing part of the stock.

### 12.2.2.1 Winter survey during 3-21. January.

The survey area was in Denmark Strait and the slope off western and north Iceland from about $27^{\circ} \mathrm{W}$ to $12^{\circ} 30^{\prime} \mathrm{W}$ and from $66^{\circ} 15^{\prime} \mathrm{N}$ to $68^{\circ} \mathrm{N}$ (Bardarson and Jonsson 2016b). The survey was initiated by scouting of 5 vessels on transects with 20 nautical mile intervals. Followed by acoustic measurement from Kolbeinseyridge to Denmark Strait giving low quality acoustic data due to wind and rough seas. During 13-20. January the whole area eastwards from Denmark strait to Bakkafloadjup, northeast of Iceland was measured in calm seas and exceptionally good conditions that provided optimal conditions for collection of high quality acoustic data of the maturing part of the stock.

Main abundance of capelin was observed from the Denmark Strait to about 20 nautical mile west of Kolbeinsey ridge, while more scattered schools were found east of the ridge. Immature capelin estimated as 9.4 billion individuals dominated in the western part of Denmark Strait, but was mixed with mature capelin further east and along the shelf edge of the Westjords peninsula. Mature capelin predominated deep off Kögurgrunn shelf area and eastwards along the continental shelf edge off north- and north-east-Iceland. In this survey about 675 thousand $t$ of mature capelin was measured. Direction and speed of capelin migration was not quantified. Simultaneous eastward migration of the capelin might have led to overestimate in stock size.

As the autumn survey used for calculating the intermediate TAC had limited coverage of the maturing stock the final TAC was based only on this winter survey. On the basis of this estimate of the mature stock and catch taken between autumn and winter survey
the Marine Research Institute recommended a TAC of 173300 t for the fishing season 2015/2016. This recommendation was in accordance with existing HCR established by WKICE (ICES, 2015).

### 12.2.2.2 Winter survey during 1-15. February

Although the first winter survey was conducted under optimal conditions, the Icelandic fishing industry offered to fund a second winter survey.

Acoustic measurements were made by two research vessels assisted by scouting of one Greenlandic fishing vessel. The survey area extended from Denmark Strait eastwards along the shelf break off north Iceland and southward along the shelf break off east Iceland (Bardarson and Jonsson 2016c). Also, shallower shelf areas were scouted.

Both research vessels left harbour on 1st February. Arni Fridriksson started acoustic measurements the 2nd February eastward from Vikurall the most western part of the survey area, but Bjarni Saemundsson started the 3d February northward from Seydisfjardardypi furthest to southeast. During $5^{\text {th }}-7$ th of February strong winds and heavy seas forced Arni Fridriksson to stop measurements and head to Isafjordur for shelter, also Bjarni Saemundsson had to stop measurements and seek shelter in Thistilfjordur from the evening of 4th February until noon of 5th February. The 9th of February the research vessels met by parallel transects off Strandagrunn giving in total one coverage of the shelf edge survey area. Given scarce observations of capelin west of Kolbeinsey ridge, it was decided to measure again the area east of Iceland where the most capelin abundance had been observed, aiming at more precise measurements with denser transects by two research vessels assisted by one scouting fishing vessel. The second coverage of east Iceland started by both research vessels furthest to the south in the survey area (following information from fishing vessels in the area) on the morning of 10th February with 5 nautical mile distance between transects where each vessel sailed every second transect. Weather conditions became difficult early on and in several occurances during this coverage research vessels had to stop measurements for several hours each time. Capelin observations in shallower areas lead to extension of coverage into the shallower area in the vicinity of Nordfjardardjup and scouting within shallow areas further north. The 12th of February, Bjarni Saemundson finished participation due to other projects, but Arni Fridriksson and Polar Amaroq continued further to north with more distant transects as capelin observations decreased.

Given the limitations of coverage due to weather conditions this survey estimate was not used for TAC advice.

### 12.3 The fishery (fleet composition, behaviour and catch)

An initial catch quota of 54000 t was recommended for the 2015/2016 fishing season, but 300000 t was agreed upon. This was changed to an intermediate quota of 44000 t in autumn 2016, and then updated to a final quota of 173300 t in winter 2017. 174 thousand t were caught in total in the 2015/2016 fishing season.

In summer 2015 more than two weeks of search for capelin by Danish vessels resulted in 0 t catch. Further, in autumn 2015, a Greenlandic vessel caught 900 t in Denmark Strait.

The distribution of the winter catches, based on logbooks for the Icelandic fleet, is shown in Figure 12.3.1. The beginning of the 2016 winter fishery had a slow start due to a low intermediate TAC based on the autumn survey. Most vessels therefore waited for results from the January survey. During last week of January and first 3 weeks of

February the Norwegian, Greenlandic and Faroese fleets caught the bulk of their quota, mainly east of Iceland, where at times the capelin became available for seine fishing although often restricted by weather.
In week 8 the Icelandic fleet was fishing from schools of capelin close to shore south off Iceland. This migration moved westward and was followed by the fishery south of the coast of west-Iceland and ended west of Breidafjordur in weeks 11-12. This spawning migration seemed not to be composed of large or dense schools. The total landings of the 2015/2016 fishing season amounted 174000 t (preliminary information), which is far below the average of catches since the beginning of the fishery.

The total annual catch of capelin in the Iceland-East Greenland-Jan Mayen area since 1964 is given by weight, season, and fleet in Table 12.3.1 and Figure 12.3.2.

Sampling from commercial catches is not considered to be adequate. 19 samples from Icelandic and Greenlandic vessels have been analysed by MRI in Iceland (length measured and age read), although samples from Norway and Faroes have not yet been processed.

The total catches in numbers by age during the summer/autumn since 1985 are given in Table 12.3.2 and for the winter since 1986 in Table 12.3.3. Similar age distribution was observed in the catches 2016 as in the survey in January 2016.

Preliminary and final TAC as well as landings for the fishing seasons since 1992/93 are given in Table 12.3.4.

### 12.4 Biological data

### 12.4.1 Growth

Seasonal growth pattern, with considerably increased growth rate during summer and autumn has been observed in this capelin stock in a study of the period 1979-1992. Where immature fish had slower growth during winter, the maturing fish had faster summer growth that continued throughout the winter until spawning in March/April, followed by almost $100 \%$ spawning mortality (Vilhjalmsson, 1994). Further, examination of the growth of immature capelin at age 1 in autumn to mature at age 2 in autumn the year after in the period 1979-2013 showed on average almost 4 fold weight increase during one year (Gudmundsdottir and Thorsteinsson, WD in 2014). This considerable weight increase and seasonal pattern in growth the year before spawning should be taken into account when deciding the timing of the capelin fisheries.
Seasonal variation of fat content is also observed. During the summer period, the fat content rises from approximately $5 \%-20 \%$ in late autumn before spawning (Engilbertsson et. al. 2012). In the fall and winter the fat content slowly declines, until the spawning migration begins in early January where the fat content drops drastically from about $15 \%-5 \%$ in mid-April. Immature capelin has much lower fat content, usually less than $3-4 \%$.

### 12.5 Methods

The objective of the HCR for the stock is to leave at least 150000 t (=Blim) for spawning (escapement strategy). The initial (preliminary), intermediate and final TACs are based on acoustic surveys.
a) The initial TAC for the coming fishing season is advised in May based autumn survey abundance estimate of immature 1 and 2 year old capelin.
b) The intermediate TAC is advised in autumn based on the biomass estimate of maturing capelin.
c) The final TAC is advised in January/February based on the biomass estimate of maturing capelin.
The initial (preliminary) quota follows a simple forecast that is based on the relation between historic observations of age 1 and 2 juvenile abundance from the acoustic autumn surveys and the corresponding final TACs nearly $11 / 2$ year later. This was done in ICES NWWG 2016 to set the initial quota for the fishing season 2016/17. Figure 12.8.1 shows this relation and the associated precautionary initial quota (blue line).
The intermediate and final TACs are set so that there is at least $95 \%$ probability that there will be 150000 t (= $\mathrm{Blim}_{\mathrm{lim}}$ ) of mature capelin left for spawning at the spawning time ( 15 march). This was done for the first time in 2015/2016 by the Icelandic Marine Research Institute and was not evaluated by ICES.

These methods were endorsed by the benchmark working group WKICE in 2015. See WKICE (ICES, 2015) and the Stock Annex for the capelin in the Iceland-East GreenlandJan Mayen area.

Previously, (since early 1980s) the stock has been managed according to an escapement strategy, leaving 400 thousand $t$ to spawning (uncertainty of the estimates were not considered). To predict the TAC for the next fishing season a model was developed early 1990s. These models were not endorsed by the benchmark working group WKSHORT 2009.

### 12.6 Reference points

During WKICE (ICES, 2015) Blim of 150000 t was defined. No other reference points are defined for this stock.

### 12.7 State of the stock

It was estimated that 304 thousand $t$ were left for spawning in spring 2016 (Table 12.7.1).

Acoustic estimation of the immature part of the stock was difficult due to adverse weather and ice conditions. The results indicate very low abundance of immature capelin.

### 12.8 Short-term forecast

The acoustic estimate of immature capelin at age 1 and 2 from the autumn survey in September 2015 was 6.2 billion which is considered a minimum estimate due to the incomplete coverage and adverse weather and ice conditions. The estimate is well below the trigger value of 50 billion and the advice according to the HCR is therefore no capelin fishery in the fishing season 2016/17 (Figure 12.8.1).

### 12.9 Uncertainties in assessment and forecast

The uncertainty of the assessment and forecast depends largely on the quality of the acoustic surveys in terms of coverage, conditions for acoustic measurements and the variance of aggregation of the capelin.

The uncertainty, mainly deriving from the variance of aggregation of the capelin, is estimated by bootstrapping (see stock annex). The CV for the immature abundance
was estimated 0.19 in the 2015 autumn survey. The CV for the mature biomass was estimated 0.26 in the 2015 autumn survey and 0.16 in the 2016 winter survey.

The autumn survey had an incomplete coverage of both the immature and maturing stock components due to drift ice and weather. Furthermore, the winter survey only measured the capelin in an eastwards direction. Simultaneous eastward migration of the capelin might have led to an overestimation of the stock size

### 12.10 Comparison with previous assessment and forecast

For the fishing season 2015/2016 an initial quota of 54000 t were advised, the intermediate TAC was 44000 t (Gudmundsdottir et. al. 2016) and the final TAC was set to 173 300 t. (Bardarson et. al. 2016). The landings were 174000 t .

### 12.11 Management plans and evaluations

See section 12.5.

### 12.12 Management considerations

The fishing season for capelin has since 1975 started in the period from late June to July/August when surveys on the juvenile part of the stock the year before have resulted in the setting of an initial (preliminary) catch quota. During summer, the availability of plankton is at its highest and the fishable stock of capelin is feeding very actively over large areas north of Iceland between Greenland and Jan Mayen, increasing rapidly in size, weight and fatness. By late September/beginning of October this period of rapid growth is over. The growth is fasted the first two years, but the weight increase is most in the year before spawning.

Timing of fishery should consider that; taken into account the large weight increase in the summer before spawning (section 12.4) it is likely that there will be more biomass of maturing fish in autumn than before the summer, although predation during the summer imposes a biomass loss. A calculation of the scale and timing of this biomass development through the fishing season cannot be done before new consumption estimates are provided. This is also supported by information for the Barents Sea capelin where it has been shown that fishing during autumn would maximize the yield, but from the ecosystem point of view a winter fishery were preferable (Gjøsæter et.al., 2002). As the biology and role in the ecosystem of these two capelin stocks are similar, this is considered to be also valid for the Icelandic capelin.

During the autumn surveys juveniles and adult capelin are often found together. This should be considered during summer fishing because the survival rate of juvenile capelin that escape through the trawlnet is unknown.

### 12.13 Ecosystem considerations

Capelin is an important forage fish and its dynamics are expected to have implications on the productivity of their predators (see further in section 7.3).

The importance of capelin in East Greenlandic waters remains to be investigated.
In Icelandic waters, capelin is the main single item in the diet of Icelandic cod, a key prey to several species of marine mammals and seabirds and also important as food for several other commercial fish species (see e.g. Vilhjálmsson, 2002).

### 12.14 Regulations and their effects

Over the years the fishery has been closed during April - late June and the season has started in July/August or later, depending on the state of the stock.
Areas with high abundances of juvenile age 1 and 2 capelin (on the shelf region off $\mathrm{NW}-\mathrm{N}$ - and NE-Iceland) have usually been closed to the summer and autumn fishery.
It is permissible to transfer catches from the purse-seine of one vessel to another vessel, in order to avoid slippage. However, if the catches are beyond the carrying capacity of the vessel and no other vessel is nearby, slippage is allowed. In recent years, reporting of such slippage has not been frequent. Industrial trawlers do not have the permission to slip capelin in order to harmonize catches to the processing.
In Icelandic waters, fishing with pelagic trawl is only allowed in limited area off the NE-coast (fishing in January) to protect juvenile capelin and to reduce the risk of affecting the spawning migration route.

A regulation calling for immediate, temporary area closures when high abundance of juveniles are measured in the catch (more than $20 \%$ of the catch composed of fish less than 14 cm ) is enforced in Icelandic waters, using on-board observers.

### 12.15 Changes in fishing technology and fishing patterns

Variable amount of the catches have been taken with pelagic trawl through the fishing seasons. Total landings in 2015/16 amounted 174 kt (preliminary numbers) ( $93 \%$ purseseine, $7 \%$ pelagic trawl). Discards are considered negligible.

### 12.16 Changes in the environment

Icelandic waters are characterized by highly variable hydrographical conditions, with temperatures and salinities depending on the strength of Atlantic inflow through the Denmark Strait and the variable flow of polar water from the north. Since 1996 the quarterly monitoring of environmental conditions of Icelandic waters shows a rise in sea temperatures north and east of Iceland, which probably also reaches farther north and northwest, as well as on the spawning grounds at South- and Southwest Iceland. It has been put forward in the 2000s that this temperature increase, may have led to a spatial shift in spawning and nursery areas (Vilhjálmsson, 2007). The acoustic surveys in autumn 2010, 2012-2015 partly confirmed the change in distribution of immatures and maturing capelin. Fisheries data suggests that the major part of the spawning still takes place on the usual grounds.

More detailed environmental description is in section 7.3.

### 12.17 Recommendations

In coming years when experience of the new HCR will be gained it is recommended that assumptions and practical operation of the HCR will be evaluated. E.g. predation/pray relationships and how SSB estimates from autumn and winter surveys should be weighted when final TAC is defined.

Studies of optimal harvesting of capelin should be conducted. These estimates should take account of growth, mortality and gear selection in relation to the timing of the fishery. Furthermore, should the role of capelin in the ecosystem be quantified to allow for a recommendation on the biomass of capelin that should be reserved for predators and whether the population size and growth of capelin predators shows a response to changes in capelin abundance.

The assessment and advice on the final TAC for capelin based on the autumn and winter surveys are issued directly to the Coastal States by the Icelandic Marine Research Institute. This process is not internationally peer reviewed prior to the release of the advice. Among the reasons for using this process is the need for fast advice once the survey result is available. The ICES ACOM procedure is more time consuming. NWWG therefore recommends that a fast track workflow based on online meetings is established if possible.

When considering the effort of research surveys for assessment of the capelin stock in 2015/2016 the initial allocation of enough effort in terms of ship time, number of ships and manpower, is recommended such that full coverage in first attempt is more likely both during autumn and winter surveys, given the demanding weather and ice conditions.

### 12.18 References

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### 12.19Tables

Table 12.2.1 Capelin. Acoustic assessment of capelin in the Iceland/Greenland/Jan Mayen area, by r/v Arni Fridriksson 16/9-4/10 2015 (Numbers in millions, biomass in tonnes).


Table 12.2.2. Icelandic Capelin. Abundance of age classes in numbers ( $10^{9}$ ) measured in acoustic surveys in autumn.

| Year | Mon | Day | AGE1 | AGE1 | AGE2 | AGE2MAT | AGE3 | AGE3MAT | AGE4 | AGE5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | IMM | MAT | IMM | MAT | IMM | MAT | MAT | MAT |
| 1978 | 10 | 16 |  |  |  | 60.0 |  | 13.9 | 0.4 |  |
| 1979 | 10 | 14 | 10.0 |  |  | 49.7 |  | 9.1 | 0.4 |  |
| 1980 | 10 | 11 | 23.5 |  |  | 19.5 |  | 4.8 |  |  |
| 1981 | 11 | 26 | 21.0 |  | 1.1 | 11.9 |  | 0.6 |  |  |
| 1982 | 10 | 2 | 68.0 |  | 1.7 | 15.0 |  | 1.6 |  |  |
| 1983 | 10 | 3 | 44.1 |  | 8.2 | 58.6 |  | 5.6 | 0.1 |  |
| 1984 | 11 | 1 | 73.8 |  | 4.6 | 31.9 |  | 10.3 | 0.3 |  |
| 1985 | 10 | 8 | 33.8 |  | 12.6 | 43.7 |  | 14.4 | 0.4 | 0.1 |
| 1986 | 10 | 4 | 58.6 |  | 1.4 | 19.9 |  | 29.8 | 0.3 |  |
| 1987 | 11 | 18 | 21.3 |  | 2.5 | 52.0 |  | 13.5 |  |  |
| 1988 | 10 | 6 | 43.9 |  | 6.7 | 53.0 |  | 17.0 | 0.4 |  |
| 1989 | 10 | 26 | 29.2 |  | 1.8 | 2.9 |  | 0.6 |  |  |
| 1990 | 11 | 8 | 24.9 |  | 1.3 | 16.4 |  | 2.7 | 0.1 |  |
| 1991 | 11 | 15 | 60.0 |  | 5.3 | 44.7 |  | 4.2 |  |  |
| 1992 | 10 | 13 | 104.6 |  | 2.3 | 54.5 |  | 4.3 | 0.1 |  |
| 1993 | 11 | 18 | 100.4 |  | 9.8 | 55.1 |  | 4.9 |  |  |
| 1994 | 11 | 25 | 119.0 |  | 6.9 | 29.2 |  | 4.4 |  |  |
| 1995 | 11 | 30 | 165.0 |  | 30.1 | 84.6 |  | 7.0 |  |  |
| 1996 | 11 | 27 | 111.9 |  | 16.4 | 70.0 |  | 15.9 |  |  |
| 1997 | 11 | 1 | 66.8 |  | 30.8 | 52.5 |  | 8.5 |  |  |
| 1998 | 11 | 13 | 121.0 |  | 5.9 | 20.5 |  | 3.3 |  |  |
| 1999 | 11 | 15 | 89.8 |  | 4.4 | 18.1 |  | 0.9 |  |  |
| 2000 | 11 | 10 | 103.7 |  | 10.9 | 11.6 | 0.1 | 0.6 |  |  |
| 2001 | 11 | 12 | 101.8 |  | 2.4 | 22.1 | 0.0 | 0.7 |  |  |
| 2002 | 11 | 12 | 1.0 |  | 0.5 |  |  |  |  |  |
| 2003 | 11 | 6 | 4.9 |  | 3.1 | 1.7 | 0.1 | 0.2 |  |  |
| 2004 | 11 | 22 | 7.9 |  | 0.1 | 7.3 |  | 0.8 | 0.0 |  |
| 2005 | 11 |  |  |  |  |  |  |  |  |  |
| 2006 | 11 | 6 | 44.7 |  | 0.3 | 5.2 |  | 0.4 |  |  |
| 2007 | 11 | 7 | 5.7 |  | 0.1 | 1.3 |  | 0.0 |  |  |
| 2008 | 11 | 17 | 7.5 | 5.1 | 0.4 | 12.1 |  | 1.8 |  |  |
| 2009 | 11 | 24 | 13.0 | 2.4 |  | 5.0 |  | 0.7 |  |  |
| 2010 | 10 | 1 | 91.6 | 9.6 | 6.3 | 25.8 | 0.1 | 0.8 | 0.02 |  |
| 2011 | 11 | 29 | 9.0 | 0.6 | 3.6 | 19.9 | 0.05 | 2.1 |  |  |
| 2012 | 10 | 3 | 18.5 | 0.9 | 2.0 | 21.2 | 0.07 | 11.4 | 0.1 |  |
| 2013 | 9 | 17 | 60.1 | 0.6 | 6.9 | 25.0 | 1.3 | 6.9 | 0.1 |  |
| 2014 | 9 | 16 | 57.0 | 1.0 | 3.3 | 26.5 | 0.2 | 7.6 | 0.1 |  |
| 2015 | 9 | 16 | 5.0 | 0.4 | 1.2 | 21.2 |  | 6.7 |  |  |

1987 - The number-at-age 1 was from survey earlier in autumn.
2005 - Scouting vessels searched for capelin. r/s ÁF measured. No samples taken for age determination. Estimated to be $<50$ thous. tonnes.
2011-Only limited coverage of the traditional capelin distribution area.
2001-2009 and 2016 - Not full coverage of stock.

Table 12.2.3. Icelandic Capelin. Mean weight (g) of age classes measured in acoustic surveys in autumn. (imm=immature, mat=mature). See footnotes in table 12.2.2.

| Year | Mon. | AGE 1 | Age 1 | AGE2 | AGE2 | AGE3 | AGE3 | AGE4 | AGE5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Імм. | MAT. | Imm. | MAT. | Imm. | MAT. | MAT | MAT |
| 1978 | 10 |  |  |  | 19.8 |  | 25.4 | 26.3 |  |
| 1979 | 10 | 6.2 |  |  | 15.7 |  | 23.0 | 20.8 |  |
| 1980 | 10 | 7.3 |  |  | 19.4 |  | 26.7 |  |  |
| 1981 | 11 | 3.6 |  | 12.3 | 19.4 |  | 22.5 |  |  |
| 1982 | 10 | 3.8 |  | 8.5 | 16.5 |  | 24.1 |  |  |
| 1983 | 10 | 5.1 |  | 9.5 | 16.8 |  | 22.5 | 23.0 |  |
| 1984 | 11 | 2.9 |  | 8.3 | 15.8 |  | 25.7 | 23.2 |  |
| 1985 | 10 | 3.8 |  | 8.5 | 15.5 |  | 23.8 | 29.5 | 31.0 |
| 1986 | 10 | 4.0 |  | 6.1 | 18.1 |  | 24.1 | 28.8 |  |
| 1987 | 11 | 2.8 |  | 8.7 | 17.9 |  | 25.8 |  |  |
| 1988 | 10 | 3.0 |  | 8.0 | 15.4 |  | 23.4 | 20.9 |  |
| 1989 | 10 | 3.5 |  | 8.0 | 12.9 |  | 24.0 |  |  |
| 1990 | 11 | 3.9 |  | 8.4 | 18.0 |  | 25.5 | 36.0 |  |
| 1991 | 11 | 4.7 |  | 7.9 | 16.3 |  | 25.4 |  |  |
| 1992 | 10 | 3.7 |  | 8.6 | 16.5 |  | 22.6 | 22.0 |  |
| 1993 | 11 | 3.6 |  | 8.9 | 16.2 |  | 23.3 |  |  |
| 1994 | 11 | 3.3 |  | 7.9 | 15.9 |  | 23.6 |  |  |
| 1995 | 11 | 3.7 |  | 7.0 | 14.0 |  | 20.8 |  |  |
| 1996 | 11 | 3.1 |  | 7.4 | 15.8 |  | 20.6 |  |  |
| 1997 | 11 | 3.3 |  | 8.5 | 14.3 |  | 20.1 |  |  |
| 1998 | 11 | 3.5 |  | 9.9 | 13.7 |  | 18.8 |  |  |
| 1999 | 11 | 3.6 |  | 8.0 | 15.4 |  | 19.5 |  |  |
| 2000 | 11 | 3.9 |  | 8.5 | 13.4 | 13.0 | 20.8 |  |  |
| 2001 | 11 | 3.8 |  | 8.8 | 16.3 | 15.7 | 23.9 |  |  |
| 2002 | 11 |  |  |  |  |  |  |  |  |
| 2003 | 11 | 7.2 |  | 14.9 | 17.0 | 22.6 | 23.7 |  |  |
| 2004 | 11 | 7.4 |  | 7.6 | 16.0 |  | 18.0 | 14.5 |  |
| 2005 |  |  |  |  |  |  |  |  |  |
| 2006 | 11 | 3.7 |  | 7.9 | 15.0 |  | 16.7 |  |  |
| 2007 | 11 | 5.5 |  | 8.6 | 14.9 |  | 15.8 |  |  |
| 2008 | 11 | 6.2 | 11.0 | 6.9 | 18.6 |  | 22.4 |  |  |
| 2009 | 11 | 5.1 | 9.8 |  | 20.0 |  | 23.8 |  |  |
| 2010 | 10 | 5.8 | 12.9 | 12.2 | 19.0 | 12.9 | 24.0 | 21.2 |  |
| 2011 | 11 | 6.8 | 11.4 | 11.1 | 18.7 | 15.8 | 24.4 |  |  |
| 2012 | 10 | 6.5 | 16.0 | 15.3 | 22.0 | 22.4 | 28.0 | 26.6 |  |
| 2013 | 9 | 5.8 | 12.6 | 10.9 | 18.0 | 11.2 | 20.9 | 23.6 |  |
| 2014 | 9 | 4.2 | 9.9 | 12.7 | 18.3 | 16.6 | 21.2 | 25.0 |  |
| 2015 | 9 | 8.5 | 12.3 | 13.4 | 18.4 | 21.5 | 23.1 |  |  |

Table 12.2.4. Icelandic Capelin. Assessment of mature capelin in the Iceland/EastGreenland/Jan Mayen area, by r/v Arni Fridriksson in January 2015 (Numbers in millions, biomass in tonnes).

|  | Length (cm) | Numbers at Age (109) |  |  |  | Numbers$\left(10^{9}\right)$ | Biomass$\left(10^{3} \mathrm{t}\right)$ | Mean weight (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 2 | 3 | 4 | 5 |  |  |  |
|  | 10 | 0.02 | 0.00 | 0.00 | 0.00 | 0.02 | 0.08 | 3.8 |
|  | 10.5 | 0.03 | 0.00 | 0.00 | 0.00 | 0.03 | 0.14 | 4.0 |
|  | 11 | 0.11 | 0.00 | 0.00 | 0.00 | 0.11 | 0.53 | 4.7 |
|  | 11.5 | 0.71 | 0.04 | 0.00 | 0.00 | 0.75 | 4.01 | 5.3 |
|  | 12 | 1.06 | 0.02 | 0.00 | 0.00 | 1.08 | 6.61 | 6.1 |
|  | 12.5 | 1.55 | 0.13 | 0.00 | 0.00 | 1.68 | 12.35 | 7.4 |
|  | 13 | 1.84 | 0.15 | 0.00 | 0.00 | 2.00 | 16.81 | 8.4 |
|  | 13.5 | 1.20 | 0.27 | 0.00 | 0.00 | 1.47 | 14.42 | 9.8 |
|  | 14 | 0.74 | 0.89 | 0.00 | 0.00 | 1.63 | 18.55 | 11.4 |
|  | 14.5 | 0.29 | 1.79 | 0.13 | 0.00 | 2.22 | 29.20 | 13.2 |
|  | 15 | 0.09 | 3.11 | 0.23 | 0.00 | 3.43 | 51.01 | 14.9 |
|  | 15.5 | 0.00 | 4.64 | 0.64 | 0.07 | 5.34 | 90.67 | 17.0 |
|  | 16 | 0.02 | 4.28 | 1.01 | 0.01 | 5.33 | 101.36 | 19.0 |
|  | 16.5 | 0.01 | 5.11 | 1.61 | 0.00 | 6.73 | 141.87 | 21.1 |
|  | 17 | 0.01 | 2.80 | 1.61 | 0.04 | 4.47 | 107.16 | 24.0 |
|  | 17.5 | 0.04 | 1.91 | 1.69 | 0.04 | 3.68 | 98.54 | 26.8 |
|  | 18 | 0.00 | 0.47 | 1.20 | 0.07 | 1.74 | 52.07 | 29.9 |
|  | 18.5 | 0.00 | 0.03 | 0.37 | 0.00 | 0.40 | 12.93 | 32.4 |
|  | 19 | 0.00 | 0.00 | 0.02 | 0.00 | 0.02 | 0.82 | 35.8 |
|  | 19.5 | 0.00 | 0.00 | 0.03 | 0.00 | 0.03 | 1.19 | 35.8 |
| TSN (109) |  | 7.75 | 25.65 | 8.55 | 0.23 | 42.18 |  |  |
| TSB ( $10^{3} \mathrm{t}$ ) |  | 66 | 486 | 203 | 6 |  | 760.34 |  |
| Mean W (g) |  | 8.5 | 19.0 | 23.7 | 24.0 |  |  | 18.0 |
| Mean L (cm) | 15.5 | 12.9 | 15.9 | 16.9 | 16.9 |  |  |  |
| \%TSN |  | 18.4 | 60.8 | 20.3 | 0.5 |  |  |  |
| SSN (109) |  | 0.8 | 23.3 | 8.4 | 0.2 | 32.8 |  |  |
| SSB ( $10^{3} \mathrm{t}$ ) |  | 11.3 | 456.5 | 201.4 | 5.6 |  | 674.8 |  |
| SMean W (g) |  | 13.4 | 19.6 | 23.9 | 24.0 |  |  | 20.6 |
| SMean L <br> (cm) | 16.2 | 14.3 | 16.0 | 16.9 | 16.9 |  |  |  |
| \%SSN |  | 2.6 | 71.0 | 25.7 | 0.7 |  |  |  |
| ISN (109) |  | 6.9 | 2.4 | 0.1 | 9.4 | 9.4 |  |  |
| ISB ( $10^{3} \mathrm{t}$ ) |  | 54.2 | 29.9 | 1.5 | 0.0 |  | 85.6 |  |
| IMean W (g) |  | 7.8 | 12.6 | 13.6 | 0.0 |  |  | 9.1 |
| IMean L (cm) | 13.2 | 12.7 | 14.5 | 14.8 | 0.0 |  |  |  |
| \%ISN |  | 73.6 | 25.2 | 1.2 | 0.0 |  |  |  |

Table 12．3．1 Capelin．The international catch since 1964 （thousand tonnes）．

|  |  | Winter season |  |  | Summer and autumn season |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\underset{\underset{\sim}{\underset{\sim}{*}}}{\stackrel{\sim}{4}}$ | $\begin{aligned} & \text { 号 } \\ & \text { تِ } \end{aligned}$ | $\begin{aligned} & \text { ̌ } \\ & \text { z } \\ & \text { o } \\ & \text { Z } \end{aligned}$ | $\begin{aligned} & \stackrel{4}{0} \\ & \stackrel{区}{4} \end{aligned}$ | $$ |  | $\begin{aligned} & \text { 号 } \\ & \underline{山 己 ~} \end{aligned}$ | $\begin{aligned} & \text { خ } \\ & \text { z } \\ & \text { Z } \end{aligned}$ |  |  | ］ |  | $\begin{gathered} \stackrel{\rightharpoonup}{\llcorner } \\ \stackrel{\rightharpoonup}{\circ} \end{gathered}$ |
| 1964 | 8.6 | － | － |  | 8.6 | － | － | － |  | － | － | 8.6 |
| 1965 | 49.7 | － | － |  | 49.7 | － | － | － |  | － | － | 49.7 |
| 1966 | 124.5 | － | － |  | 124.5 | － | － | － |  | － | － | 124.5 |
| 1967 | 97.2 | － | － |  | 97.2 | － | － | － |  | － | － | 97.2 |
| 1968 | 78.1 | － | － |  | 78.1 | － | － | － |  | － | － | 78.1 |
| 1969 | 170.6 | － | － |  | 170.6 | － | － | － |  | － | － | 170.6 |
| 1970 | 190.8 | － | － |  | 190.8 | － | － | － |  | － | － | 190.8 |
| 1971 | 182.9 | － | － |  | 182.9 | － | － | － |  | － | － | 182.9 |
| 1972 | 276.5 | － | － |  | 276.5 |  | － | － |  | － | － | 276.5 |
| 1973 | 440.9 | － | － |  | 440.9 | － | － | － |  | － | － | 440.9 |
| 1974 | 461.9 | － | － |  | 461.9 | － | － | － |  | － | － | 461.9 |
| 1975 | 457.1 | － | － |  | 457.1 | 3.1 | － | － |  | － | 3.1 | 460.2 |
| 1976 | 338.7 | － | － |  | 338.7 | 114.4 | － | － |  | － | 114.4 | 453.1 |
| 1977 | 549.2 | － | 24.3 |  | 573.5 | 259.7 | － | － |  | － | 259.7 | 833.2 |
| 1978 | 468.4 | － | 36.2 |  | 504.6 | 497.5 | 154.1 | 3.4 |  | － | 655.0 | 1，159．6 |
| 1979 | 521.7 | － | 18.2 |  | 539.9 | 442.0 | 124.0 | 22.0 |  | － | 588.0 | 1，127．9 |
| 1980 | 392.1 | － | － |  | 392.1 | 367.4 | 118.7 | 24.2 |  | 17.3 | 527.6 | 919.7 |
| 1981 | 156.0 | － | － |  | 156.0 | 484.6 | 91.4 | 16.2 |  | 20.8 | 613.0 | 769.0 |
| 1982 | 13.2 | － | － |  | 13.2 | － | － | － |  | － | － | 13.2 |
| 1983 | － | － | － |  | － | 133.4 | － | － |  | － | 133.4 | 133.4 |
| 1984 | 439.6 | － | － |  | 439.6 | 425.2 | 104.6 | 10.2 |  | 8.5 | 548.5 | 988.1 |
| 1985 | 348.5 | － | － |  | 348.5 | 644.8 | 193.0 | 65.9 |  | 16.0 | 919.7 | 1，268．2 |
| 1986 | 341.8 | 50.0 | － |  | 391.8 | 552.5 | 149.7 | 65.4 |  | 5.3 | 772.9 | 1，164．7 |
| 1987 | 500.6 | 59.9 | － |  | 560.5 | 311.3 | 82.1 | 65.2 |  | － | 458.6 | 1，019．1 |
| 1988 | 600.6 | 56.6 | － |  | 657.2 | 311.4 | 11.5 | 48.5 |  | － | 371.4 | 1，028．6 |
| 1989 | 609.1 | 56.0 | － |  | 665.1 | 53.9 | 52.7 | 14.4 |  | － | 121.0 | 786，1 |
| 1990 | 612.0 | 62.5 | 12.3 |  | 686.8 | 83.7 | 21.9 | 5.6 |  | － | 111.2 | 798.0 |
| 1991 | 202.4 | － | － |  | 202.4 | 56.0 | － | － |  | － | 56.0 | 258.4 |
| 1992 | 573.5 | 47.6 | － |  | 621.1 | 213.4 | 65.3 | 18.9 | 0.5 | － | 298.1 | 919.2 |
| 1993 | 489.1 | － | － | 0.5 | 489.6 | 450.0 | 127.5 | 23.9 | 10.2 | － | 611.6 | 1，101．2 |
| 1994 | 550.3 | 15.0 | － | 1.8 | 567.1 | 210.7 | 99.0 | 12.3 | 2.1 | － | 324.1 | 891.2 |
| 1995 | 539.4 | － | － | 0.4 | 539.8 | 175.5 | 28.0 | － | 2.2 | － | 205.7 | 745.5 |
| 1996 | 707.9 | － | 10.0 | 5.7 | 723.6 | 474.3 | 206.0 | 17.6 | 15.0 | 60.9 | 773.8 | 1，497．4 |
| 1997 | 774.9 | － | 16.1 | 6.1 | 797.1 | 536.0 | 153.6 | 20.5 | 6.5 | 47.1 | 763.6 | 1，561．5 |
| 1998 | 457.0 | － | 14.7 | 9.6 | 481.3 | 290.8 | 72.9 | 26.9 | 8.0 | 41.9 | 440.5 | 921.8 |
| 1999 | 607.8 | 14.8 | 13.8 | 22.5 | 658.9 | 83.0 | 11.4 | 6.0 | 2.0 | － | 102.4 | 761.3 |
| 2000 | 761.4 | 14.9 | 32.0 | 22.0 | 830.3 | 126.5 | 80.1 | 30.0 | 7.5 | 21.0 | 265.1 | 1，095．4 |
| 2001 | 767.2 | － | 10.0 | 29.0 | 806.2 | 150.0 | 106.0 | 12.0 | 9.0 | 17.0 | 294.0 | 1，061．2 |
| 2002 | 901.0 | － | 28.0 | 26.0 | 955.0 | 180.0 | 118.7 | － | 13.0 | 28.0 | 339.7 | 1，294．7 |
| 2003 | 585.0 | － | 40.0 | 23.0 | 648.0 | 96.5 | 78.0 | 3.5 | 2.5 | 18.0 | 198.5 | 846.5 |


*preliminary, provided by working group members.

Table 12.3.2 Icelandic capelin. The total international catch of capelin in the Iceland-East Green-land-Jan Mayen area by age group in numbers (billions) and the total catch by numbers and weight (thousand tonnes) in the autumn season (August-December) since 1985.

| Year | AGE 1 | AGE 2 | AGE 3 | AGE 4 | Age 5 | Total NUMBER | Total WEIGHT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0.8 | 25.6 | 15.4 | 0.2 |  | 42.0 | 919.7 |
| 1986 | + | 10.0 | 23.3 | 0.5 |  | 33.8 | 772.9 |
| 1987 | + | 27.7 | 6.7 | + |  | 34.4 | 458.6 |
| 1988 | 0.3 | 13.6 | 5.4 | + |  | 19.3 | 371.4 |
| 1989 | 1.7 | 6.0 | 1.5 | + |  | 9.2 | 121.0 |
| 1990 | 0.8 | 5.9 | 1.0 | + |  | 7.7 | 111.2 |
| 1991 | 0.3 | 2.7 | 0.4 | + |  | 3.4 | 56.0 |
| 1992 | 1.7 | 14.0 | 2.1 | + |  | 17.8 | 298.1 |
| 1993 | 0.2 | 24.9 | 5.4 | 0.2 |  | 30.7 | 611.6 |
| 1994 | 0.6 | 15.0 | 2.8 | + |  | 18.4 | 324.1 |
| 1995 | 1.5 | 9.7 | 1.1 | + |  | 12.3 | 205.7 |
| 1996 | 0.2 | 25.2 | 12.7 | 0.2 |  | 38.4 | 773.7 |
| 1997 | 1.8 | 33.4 | 10.2 | 0.4 |  | 45.8 | 763.6 |
| 1998 | 0.9 | 25.1 | 2.9 | + |  | 28.9 | 440.5 |
| 1999 | 0.3 | 4.7 | 0.7 | + |  | 5.7 | 102.4 |
| 2000 | 0.2 | 12.9 | 3.3 | 0.1 |  | 16.5 | 265.1 |
| 2001 | + | 17.6 | 1.2 | + |  | 18.8 | 294.0 |
| 2002 | + | 18.3 | 2.5 | + |  | 20.8 | 339.7 |
| 2003 | 0.3 | 11.8 | 1 | + |  | 14.3 | 199.5 |
| 2004 | + | 5.3 | 0.5 | - |  | 5.8 | 92.0 |
| 2005 | - | 0.4 | + | - |  | 0.4 | 9.0 |
| 2006 | - | - | - | - |  | - | - |
| 2007 | - | - | - | - |  | - | - |
| 2008 | - | - | - | - |  | - | - |
| 2009 | - | - | - | - |  | - | - |
| 2010 | 0.01 | 0.23 | 0.02 | - |  | 0.25 | 5.4 |
| 2011 | - | 2.45 | 1.61 | - | 0.08 | 4.13 | 72.1 |
| 2012 | - | 0.2 | 0.2 | - | - | 0.4 | 10.4 |
| 2013 | - | - | - | - | - | - | - |
| 2014 | 0.01 | 2.22 | 0.6 | 0.02 | - | 2.8 | 45.5 |
| 2015 | 0.03 | 0.08 | 0.03 |  |  | 1.4 | 2.5 |

Table 12.3.3 Icelandic capelin. The total international catch of capelin in the Iceland-East Green-land-Jan Mayen area by age group in numbers (billions) and the total catch by numbers and weight (thousand tonnes) in the winter season (January-March) since 1986.

| Year | AGE 1 | AGE 2 | AGE 3 | AGE 4 | AGE 5 | Total number | Total weight |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 |  | 0.1 | 9.8 | 6.9 | 0.2 | 17.0 | 391.8 |
| 1987 |  | + | 6.9 | 15.5 | - | 22.4 | 560.5 |
| 1988 |  | + | 23.4 | 7.2 | 0.3 | 30.9 | 657.2 |
| 1989 |  | 0.1 | 22.9 | 7.8 | + | 30.8 | 665.1 |
| 1990 |  | 1.4 | 24.8 | 9.6 | 0.1 | 35.9 | 686.8 |
| 1991 |  | 0.5 | 7.4 | 1.5 | + | 9.4 | 202.4 |
| 1992 |  | 2.7 | 29.4 | 2.8 | + | 34.9 | 621.1 |
| 1993 |  | 0.2 | 20.1 | 2.5 | + | 22.8 | 489.6 |
| 1994 |  | 0.6 | 22.7 | 3.9 | + | 27.2 | 567.1 |
| 1995 |  | 1.3 | 17.6 | 5.9 | + | 24.8 | 539.8 |
| 1996 |  | 0.6 | 27.4 | 7.7 | + | 35.7 | 723.6 |
| 1997 |  | 0.9 | 29.1 | 11 | + | 41.0 | 797.6 |
| 1998 |  | 0.3 | 20.4 | 5.4 | + | 26.1 | 481.3 |
| 1999 |  | 0.5 | 31.2 | 7.5 | + | 39.2 | 658.9 |
| 2000 |  | 0.3 | 36.3 | 5.4 | + | 42.0 | 830.3 |
| 2001 |  | 0.4 | 27.9 | 6.7 | + | 35.0 | 787.2 |
| 2002 |  | 0.1 | 33.1 | 4.2 | + | 37.4 | 955.0 |
| 2003 |  | 0.1 | 32.2 | 1.9 | + | 34.4 | 648.0 |
| 2004 |  | 0.6 | 24.6 | 3 | + | 28.3 | 542.9 |
| 2005 |  | 0.1 | 31.5 | 3.1 | - | 34.7 | 692.0 |
| 2006 |  | 0.1 | 10.4 | 0.3 | - | 10.8 | 230.0 |
| 2007 |  | 0.3 | 19.5 | 0.5 | - | 20.3 | 376.8 |
| 2008 |  | 0.5 | 10.6 | 0.4 | - | 11.5 | 202.4 |
| 2009 |  | 0.1 | 0.6 | 0.1 | - | 0.7 | 15.1 |
| 2010 |  | 0.7 | 5.3 | 0.9 | 0.01 | 6.9 | 150.7 |
| 2011 |  | 0.1 | 16.2 | 0.6 | - | 17.0 | 385.2 |
| 2012 | 0.02 | 0.6 | 25.0 | 6.1 | 0.02 | 31.8 | 674.4 |
| 2013 | - | 0.3 | 12.1 | 9.7 | 0.2 | 22.3 | 541.0 |
| 2014 | - | 0.1 | 4.8 | 1.3 | + | 6.1 | 141.8 |
| 2015 | - | 0.3 | 17.5 | 4.7 | 0.1 | 22.7 | 471.9 |
| 2016 |  | 0.4 | 5.5 | 2.0 | 0.02 | 8.0 | 171.1 |

Table 12.3.4. Initial quota and final TAC by seasons.

| Fishing season | Initial advice | Final tac | Landings |
| :---: | :---: | :---: | :---: |
| 1992/931 | 500 | 900 | 788 |
| 1993/941 | 900 | 1250 | 1179 |
| 1994/95 | 950 | 850 | 842 |
| 1995/961 | 800 | 1390 | 930 |
| 1996/971 | 1100 | 1600 | 1571 |
| 1997/98 | 850 | 1265 | 1245 |
| 1998/99 | 950 | 1200 | 1100 |
| 1999/00 | 866 | 1000 | 934 |
| 2000/01 | 650 | 1090 | 1065 |
| 2001/02 | 700 | 1300 | 1249 |
| 2002/03 | 690 | 1000 | 988 |
| 2003/042 | 555 | 900 | 741 |
| 2004/053 | 335 | 985 | 783 |
| 2005/06 | No fishery | 235 | 238 |
| 2006/07 | No fishery | 385 | 377 |
| 2007/08 | 207 | 207 | 202 |
| 2008/094 | No fishery |  | 15 |
| 2009/10 | No fishery | 150 | 151 |
| 2010/11 | No fishery | 390 | 391 |
| 2011/12 | 366 | 765 | 747 |
| 2012/13 | No fishery | 570 | 551 |
| 2013/141 | No fishery | 160 | 142 |
| 2014/15 | 2255 | 580 | 517 |
| 2015/166 | No fishery5 | 173 | 174 |

1) The final TAC was set on basis of autumn surveys in the season.
2) Indices from April 2003 were projected back to October 2002.
3) The initial quota was set on a basis of an acoustic survey in June/July 2004
4) No fishery was allowed, 15000 t was assigned to scouting vessels.
5) Initial advice based on low probability of exceeding final TAC.
6) Preliminary landings.

Table 12.7.1 Icelandic capelin in the Iceland-East Greenland-Jan Mayen area since the fishing season 1978/79. (A fishing season e.g. 1978/79 starts in summer 1978 and ends in March 1979). Recruitment of 1 year old fish (unit $10^{9}$ ) is given for 1 August at the beginning of the season. Spawningstock biomass (' 000 t ) is given at the time of spawning at the end of the fishing season. Landings (' 000 t ) are the sum of the total landings in the season

| Season (Summer/winter) | Recruitment | LANDINGS | SPAWNING-STOCK BIOMASS |
| :---: | :---: | :---: | :---: |
| 1978/79 | - | 1195 | 600 |
| 1979/80 | 22 | 980 | 300 |
| 1980/81 | 23.5 | 684 | 170 |
| 1981/82 | 21 | 626 | 140 |
| 1982/83 | 68 | 0 | 260 |
| 1983/84 | 44.1 | 573 | 440 |
| 1984/85 | 73.8 | 896 | 460 |
| 1985/86 | 33.8 | 1312 | 460 |
| 1986/87 | 58.6 | 1334 | 420 |
| 1987/88 | 2.6 | 1116 | 400 |
| 1988/89 | 43.9 | 1036 | 440 |
| 1989/90 | 29.2 | 807 | 115 |
| 1990/91 | 27.2 | 313 | 330 |
| 1991/92 | 60 | 677 | 475 |
| 1992/93 | 104.6 | 788 | 499 |
| 1993/94 | 100.4 | 1178 | 460 |
| 1994/95 | 119 | 864 | 420 |
| 1995/96 | 165 | 930 | 830 |
| 1996/97 | 111.9 | 1570 | 430 |
| 1997/98 | 66.8 | 1246 | 492 |
| 1998/99 | 121 | 1100 | 500 |
| 1999/00 | 89.8 | 932 | 650 |
| 2000/01 | 103.7 | 1071 | 450 |
| 2001/02 | 101.8 | 1249 | 475 |
| 2002/03 | - | 988 | 410 |
| 2003/04 | 4.9 | 742 | 535 |
| 2004/05 | 7.9 | 784 | 602 |
| 2005/06 | - | 247 | 400 |
| 2006/07 | 44.7 | 377 | 410 |
| 2007/08 | 5.7 | 203 | 406 |
| 2008/09 | 12.6 | 150 | 328 |
| 2009/10 | 15.4 | 151 | 410 |
| 2010/11 | 101.2 | 391 | 411 |
| 2011/12 | 9.6 | 747 | $418$ |
| 2012/13 | 19.4 | 551 | 417 |
| 2013/14 | 60.7 | 142 | 424 |
| 2014/15 | 58 | 518 | 460 |
| 2015/16 | 5.4 | $174^{* *}$ | 304* |

*Based on predation model in current HCR.
** preliminary

### 12.20Figures



Figure 12.2.1. Icelandic capelin. Cruise tracks, relative density and distribution of capelin during an acoustic survey by r/v Arni Fridriksson during 16 September-4 October 2015.


Figure 12.2.2. Icelandic capelin. Distribution of immature and maturing capelin biomass during an acoustic survey by r/v Arni Fridriksson during 16 September - 4 October 2015.


Figure 12.2.2. Icelandic capelin. Indices of immature 1 and immature 2 years old capelin from acoustic surveys in autumn since 1979.


Figure 12.2.3. Icelandic capelin. Survey tracks of r/s Arni Fridriksson during 13 - 20 January 2016.


Figure 12.3.1. Icelandic capelin. Distribution of the catches in the fishing season 2015/16 based on data from logbooks of the Icelandic fleet.


Figure 12.3.2. Icelandic capelin. The total catch (in thousand tonnes) of the Icelandic capelin since 1963/64 by season.


Figure 12.8.1 Capelin in Subareas 5 and 14 and Division 2.a west of $5^{\circ} \mathrm{W}$. Catch advice according to the proposed stochastic HCR, based on the measured number of immature capelin about 15 months earlier. The figure shows the estimated final TAC (black unbroken line) and the initial (preliminary) TAC (blue dashed line). The latter is set using a $U_{\text {trigger }}$ (red vertical line) of 50 billion immature fish, with a cap on the initial (preliminary) TAC of 400 kt . The green lines show the index value from the autumn survey 2015, with the corresponding initial TAC for 2016/2017 shown on the $y$-axis. (The figure adapted from stock-annex, WKICE 2015).

## 13 Overview on ecosystem, fisheries and their management in Greenland waters

### 13.1 Ecosystem considerations

The marine ecosystem around Greenland is located from arctic to Subarctic regions. The water masses in East Greenland are composed of the polar East Greenland Current and the warm and saline Irminger Current of Atlantic origin. As the currents round Cape Farewell at Southernmost Greenland the saline, warm Irminger water subducts the colder polar water and forms the relatively warm West Greenland Current. This flows along the West Greenland coast mixing extensively as it flows north. This current is of importance in the transport of larval and juvenile fish along the coast for important species such as cod and Greenland halibut. Additionally, cod from Icelandic waters spawning south and west of Iceland occasionally enters Greenland waters via the Irminger current and is distributed along both the Greenland East and West coast (Figure 1).


Figure 1. Spawning areas, egg and larval transport of Atlantic cod (Gadus morhus) in Greenlandic and Icelandic waters.

Depending of the relative strength of the two East Greenland currents, the Polar Current and the Irminger Current, the marine environment experience extensive variability with respect to the hydrographical properties of the West Greenland Current. The general effects of such changes have been increased production during warm periods as compared to cold periods, and resulted in extensive distribution and productivity changes of many commercial stocks. Historically, cod is the most prominent example of such a change (Holger \& Wieland 2008).

In recent years, temperature have increased significantly in Greenland waters. In West Greenland the sea temperature have increased particularly compared to the years in 1970s-mid1990s and historical highs was registered in 2005 for the time-series 18802012 (Figure 2).


Figure 2. Mean temperature on top of Fylla Bank (located outside Nuuk Fjord, 0-40 m depth) in the middle of June for the period 1950-2013. The curves are 3 year running mean values. The magenta/purple line is extended back to 1876 using Smed-data for area A1. From Ribergaard (2014).

Temperature in the centre of the Irminger Sea, in the depth interval 200-400m, shows no such clear long-term trend (ICES 2013c). However, Rudels et al. (2012) finds that between 1998-2010, the salinity and temperature of the deep water in the Greenland Sea increased. Furthermore, increasing temperatures in the Atlantic Water entering the Arctic in the Fram Strait has increased throughout the period 1996-2012, though with the highest observation in 2006 (ICES 2013c). Such environmental changes might well propagate to different trophic levels. Accordingly, shrimp biomass fluctuations in Greenland waters as a result of environmental changes could affect fish predators such as cod (Hvingel \& Kingsley 2006) and the other way around.

The primary production period in Greenland is timely displaced along the coast due to increasing sea ice cover and a shorter summer period moving north (Blicher et al. 2007), but the main primary production takes place in May-June (Figure 3). The large latitudinal gradient spanned by Greenland, the ecosystem structure shifts moving north. For instance, the secondary producer assembly (e.g. mainly copepods) shifts from being dominated by smaller Atlantic species (Calanus finmarchicus and Calanus glacialis) to being increasingly dominated by the (sub)arctic species Calanus hyperboreus.


Figure 3. Annual variation in algal biomass and productivity at the inlet of Nuuk Fjord. a: chlorophyll ( $\mu \mathrm{g} \mathrm{l}^{-1}$ ), b: fluorescence, c: primary production ( $\mathrm{mg} \mathrm{C} \mathrm{m}^{-2} \mathrm{~d}^{-1}$ ). Dots represent sampling points. From Mikkelsen et al. (2008).

Recently, the distribution of commercial species such as cod and shrimp has shifted considerably in the north. Such shifts have previously been associated with temperature, and may very well be linked to the observed increase in temperature. Additionally, changes in growth of fish may also increase as a result of temperature changes as seen for both Greenland halibut (Sünksen et al. 2008) and cod (Hovgård and Wieland 2008).

In recent years more southerly distributed species not normally seen in Greenland waters such as pearlside (Maurolicus muelleri), Whiting (Merlangius merlangus), blackbelly rosefish (Helicolenus dactylopterus), angler (Lophius piscatorius) and snake pipefish (Entelurus aequoreus) have been observed in surveys in offshore West and East Greenland and inshore West Greenland and their presence is possibly linked to increases in temperature (Møller et al. 2010).

In 2011, a mackerel (Scomber scombrus) fishery was initiated in East Greenland waters. Previous to this, no catches had ever been reported for this area and in 2013 mackerel was for the first time documented along the West Greenland coast. The reason(s) for the increased abundance of mackerel in Greenlandic waters has not been clarified, however factors such as changes in the regime for their usual food resources, a densitydependent effect and increased temperatures have been proposed (ICES 2013a). The effects of increased pelagic fish abundance and their distributional shifts on demersal fish are unknown.

### 13.1.1 Atmospheric conditions

Cod and possibly other species recruitment in Greenland waters is significantly influenced by environmental factors such as sea surface temperatures in the important Dohrn Bank region during spawning and hence by air temperatures together with the meridional wind in the region between Iceland and Greenland (Stein \& Borovkov 2004). The effect of the meridional wind component in the region off South Greenland on the first winter of the offspring appears to play a vital role for the cod recruitment process. For instance, during 2003, when the strong 2003 YC was born, negative anomalies were more than $-2.0 \mathrm{~m} / \mathrm{sec}$, and that particular YC was large in East Greenland waters. In general, it seems that during anomalous east wind conditions during summer months, anomalous numbers of 0-group cod are also found in Greenland waters.


Figure 4. NAO Index (Dec-Feb) 1950-2012.

## The NAO index

The NAO index, as given for 1950-2012 (Figure 4), shows negative values for winter (December-February) 2008/2009, 2009/2010 and 2010/2011. The 2009/2010 index is the strongest negative index (-1.64), encountered since 1950.

During the second half of the last century the 1960s were generally "low-index" years while the 1990s were "high-index" years. A major exception to this pattern occurred between the winter preceding 1995 and 1996, when the index flipped from being one of its most positive (1.36) values to a negative value ( -0.62 ). The direct influence of NAO on Nuuk winter mean air temperatures is as follows: A "low-index" year corresponds to warmer-than-normal years. Colder-than-normal temperature conditions at Nuuk are linked to "high-index" years and hence indicate a negative correlation of Nuuk winter air temperatures with the NAO. Correlation between both time-series is significant ( $\mathrm{r}=-0.73, \mathrm{p} \ll 0.001$; Stein 2004). This is seen for instance in 2009, 2010 and 2011 where air temperature anomalies at Nuuk (1.0K, 4.8 K and 2.9 K ) where associated with low NAO values (Figure 5). The 2010 air temperature anomaly (4.0K) was the highest recorded, and was associated with the largest negative NAO anomaly (see Figure 6).


Figure 5. Time-series of annual mean winter (DEC-FEB) air temperature anomalies (K) at Nuuk (1876-2012, rel. 1961-1990)


Figure 6. Time-series of annual mean air temperature anomalies (K) at Nuuk (1876-2011, rel. 19611990), and 13 year running mean.

## Zonal wind components

A negative anomaly of zonal wind components for the Northwest Atlantic is associated with atmospheric conditions in the Iceland-Greenland region enclosing strong easterly winds (Figure 7, top left panel). These winds favour surface water transports from Iceland to East Greenland and was particularly strong in 2009, while it was completely different during the same months in 2010 (Figure 7). During May-August in 2011, the cells of negative anomalies were seen to the east of Newfoundland (anomalies < 3.0 $\mathrm{m} / \mathrm{sec}$ ), and to the east of Iceland.


Figure 7. Zonal wind components for the North Atlantic (May-Aug), anomalies from 1981-2010. top left: 2009; top right: 2010; bottom left: 2011.

## Meridional wind components

As discussed in Stein and Borovkov (2004), the meridional wind component (Dec-Jan) from the Southwest Greenland region correlated positively with the trend in Greenland cod recruitment time-series (first winter of age-0 cod). During winter 2009/2010, positive meridional wind anomalies were observed Southwest Greenland (Figure 8, top left panel). During winter 2010/2011, the center of positive meridional wind anomalies had moved to the Davis Strait region (Figure 5, top right panel), and during winter 2011/2012, positive meridional wind anomalies had moved to the Northeast off Newfoundland (bottom left panel in Figure 8).


Figure 8. Meridional wind component (Dec-Jan), anomalies from 1981-2010. top left: 2009/2010; top right: 2010/2011; bottom left: 2011/2012;

### 13.1.2 Description of the fisheries

Fisheries targeting marine resources off Greenland can be divided into inshore and offshore fleets. The majority of the Greenland fleet has been built up through the 60s and is today comprised of approx. 450 larger vessels and a big fleet of small boats. It is estimated that around 1700 small boats are dissipating in some sort of artisanal fishery mainly for private use or in the poundnet fishery.

Active fishing fleet reported to Greenland statistic by GRT in 1996 - no later number is available:

| ALL FLEET (N) | $<$ 5GRT | 6-10GRT | 11-20GRT | 21-80GRT | >80GRT |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 441 | $31 \%$ | $34 \%$ | $2 \%$ | $9 \%$ | $6 \%$ |

There is a large difference between the fleet in the northern and southern part of Greenland. In south, were the cod fishery has historically been important the average vessel age is 22 years, in north only 9 years as it is mostly comprised of smaller boats targeting Greenland halibut using longlines.

### 13.1.3 Inshore fleets

The fleet is constituted by a variety of different platforms from dog sledges used for ice fishing, to small multipurpose boats engaged in whaling or deploying passive gears such as gillnets, poundnets, traps, dredges and longlines.

In the northern areas from Disko Bay at $72^{\circ} \mathrm{N}$ and north to Upernavik at $74{ }^{\circ} 30 \mathrm{~N}$, dog sledge are the platforms in winter and small open vessels the units in summer, both fishing with longlines to target Greenland halibut in the ice fjords. The main bycatch from this fishery is redfish, Greenland shark, roughhead grenadier and in recent years' cod in Disko Bay.

The coastal shrimp fisheries are distributed along most of the West coast from $61-72^{\circ} \mathrm{N}$. The main bycatch with the inshore shrimp trawlers is juvenile redfish, cod and Greenland halibut. An inshore shrimp fishery is conducted mainly in Disko Bay. Sorting grid is mandatory for the shrimp fishery; however, several small inshore shrimp trawlers have dispensation for using sorting grid.

Cod is targeted all year, but with a peak in effort in June-July as cod in this period is accessible in shallow waters facilitating the use of the main gear types, pound and gillnets. Bycatches are limited and are mainly Greenland cod (Gadus ogac) and wolffish.

In the recent years there has been an increasing exploitation rate for lumpfish. The fishing season is short, with the majority of the catch being caught in May-June. Lumpfish is caught along most of the West coast and is caught using gillnets. In small areas there is a substantial by catch of birds, especially common eiders (Somateria mollissima)

The scallop fishery is conducted with dredges at the West coast from $64-72{ }^{\circ} \mathrm{N}$, with the main landings at $66^{\circ} \mathrm{N}$. Bycatch in this fishery is considered insignificant.

Snow crabs are caught in traps in areas $62-70^{\circ} \mathrm{N}$. Problems with bycatch are at present unknown, but are believed to be insignificant.

Salmon are caught in August-October with drifting nets and gillnets. The fishery is a mix of salmon of European and North American origin.

The coastal fleets fishing for Atlantic cod, snow crab, scallops and shrimp are regulated by licenses, TAC and closed areas. Fishery for salmon and lumpfish are unregulated.

### 13.1.4 Offshore fleets

Apart from the Greenland fleet, the marine resources in Greenland waters are exploited by several nations, mainly EU, Iceland and Norway using bottom and pelagic trawls as well as longlines.

The demersal offshore fishery is comprised of vessels primarily fishing Greenland halibut, shrimp, redfish and cod. Greenland halibut and redfish have been targeted since 1985 using demersal otter board trawls with a minimum mesh size of 140 mm . A cod fishery has previously been conducted since 1920s in West Greenland offshore waters but was absent from 1992-2000s. In 2010 the cod fishery was closed off West Greenland and catches has been insignificant since. The Greenland offshore shrimp fleet consist of 15 freezer trawlers. They exclusively target shrimp stocks off West and East Greenland with landings slightly below 100000 t . The shrimp fleet is close to or above 80 BT and $75 \%$ of the fleet process the shrimp on board. Shrimp trawls are used with a minimum mesh size of 44 mm and a mandatory sorting grid ( 22 mm ) to avoid bycatch of juvenile fish. The three most economically important fish species in Greenland: Greenland halibut, redfish and cod are found in relatively small proportions in the bycatch.

However, when juvenile fish are caught, even small biomasses can correspond to relatively large numbers.

Longliners are operating on both the East and West coast with Greenland halibut and cod as targeted species. Bycatches include roundnose grenadier, roughhead grenadier, tusk, Atlantic halibut and Greenland shark (Gordon et al. 2003).

The pelagic fishery in Greenland waters is conducted in East Greenland and currently targeted species are mackerel and pelagic redfish. A relatively small fishery after herring is carried out in the border area between Greenland, Iceland and Jan Mayen. A capelin fishery has previously been done but as the Greenland share of the TAC is taken in other waters. Generally, the pelagic fishery in Greenland is very clean, with small amounts of bycatch seen.

The demersal and pelagic offshore fishing, together with longlines are managed by TAC, minimum landing sizes, gear specifications and irregularly closed areas.

### 13.2 Overview of resources

In the last century the main target species of the various fisheries in Greenland waters have changed. A large international fleet in the 50s and 60s landed large catches of cod reaching historic high in 1962 with about 450000 t . The offshore stock collapsed in the late 60s-early 70s due to heavy exploitation and possible due to environmental conditions. Since then the stock has been low, with occasional larger YC being transported from Iceland (i.e. 1984 and 2003). Since 2010, the cod biomass has been concentrated in the spawning grounds off East Greenland. Following the cod collapse, the offshore shrimp fishery started in 1969 and has been increasing up to 2003 reaching a catch level close to 150000 t . The stock decreased thereafter and is now at the low 1990 level with an advised TAC for 2015 of 60000 t . The advised TAC for 2016 increased to 90000 t .

### 13.2.1 Shrimp

The shrimp (Pandalus borealis) stock in Greenland waters has been declining since 2003. The stock in East Greenland is at a low level based on available information. The 2003 West Greenland shrimp biomass was at the highest in the time-series, but it has since decreased.

### 13.2.2 Snow crab

The biomass of snow crab (Chionoecetes opilio) in West Greenland waters has decreased substantially since 2001. Snow crab has been exploited inshore since the mid90s and offshore since 1999. Total landings have since 2010 been reported at around 2000 t a decrease from a high level in 2001 at 15000 t . After several years of decreasing cpue it now appears to have stabilized at low levels in the majority of areas.

### 13.2.3 Scallops

The status of scallops in Greenland is unknown. From the mid80s to the start 90s landings were between 4-600 t yearly, increased to around 2000 t in late 1990ies. Catches decreased again and is below 600 tons in 2014. The fishery is based on license and is exclusively at the west coast between $20-60 \mathrm{~m}$. The growth rate is considered very low reaching the minimum landing size on 65 mm in 10 years.

### 13.2.4 Squids

The status of squids in Greenland waters are unknown.

### 13.2.5 Cod

Since 2015, assessment and advice for cod in Greenland water take into account that three different stocks, based on spawning areas and genetics, are the basis for the cod fishery and the following management is therefore recommended for different three areas: a) inshore in Western Greenland (NAFO subdivision 1A-1F), b) offshore Western Greenland (NAFO subdivision 1A-1E) and offshore Eastern and South Greenland (ICES Subarea 14.b and NAFO subdivision 1F). Current landings for inshore cod are 25200 t , and have steadily increased since 2009 where landings were 7000 t . Landing from offshore Western Greenland was minor (less than 500 t since 2006) until 2015 where catches increased to 4600 t . From offshore Eastern Greenland area 2015 landing was 15800 t , an increase from the 2011-2013 level at 5000 t .

Catches are high compared to the last three decades, however they are only a fraction of the landings caught in the 1950's and 1960's. Recruitment has been negligible since the 1984 and 1985-year class, though it has improved in the last decade, especially inshore, where the 2009 YC is the best seen in the time-series since 1982. In 2007 and 2009 dense concentrations of unusual large cod were documented to be actively spawning off East Greenland, and management actions have been taken to protect these spawning aggregations. The inshore fishery has been regulated since 2009 and the offshore fishery is managed with license and minimum size $(40 \mathrm{~cm})$. As a response to the favourable environmental conditions (large shrimp stock, high temperatures) there is a possibility that the offshore cod will rebuild to historical levels if managed with this objective. A management plan with the objective of achieving this goal has been implemented for the fishing seasons 2014-2016. Several YC are present in the inshore fishery, and with the stable recruitment in recent years and widespread fishery there are several indications that the stock is experiencing favourable conditions and that recruitment is not impaired despite an increased fishing effort in later years. However in 2015 signs of increasing fishing pressure is seen as the biomass index in the inshore survey is stable and recruitment is low.

### 13.2.6 Redfish

Redfish (Sebastes mentella and Sebastes norvegicus) are primarily caught of East Greenland. Catches have been small since 1994, but recently large year classes have given rise to a significant fishery with catches in 2010-15 being around 8000 t . This includes both redfish species, but the majority (e.g. $\sim 70 \%$ ) is identified asS. mentella. Recent East Greenland survey estimates indicate a decline in S. mentella while S. norvegicus is increasing.

### 13.2.7 Greenland halibut

Greenland halibut in the Greenland area consist of at least two stocks and several components; the status of the inshore component is not known, but it has sustained catches of 15-20 000 t annually, taken primarily in the northern area (north of $68^{\circ} \mathrm{N}$ ). The offshore stock component in West Greenland (NAFO SA $0+1$ ) is a part of a shared stock between Greenland and Canada. The stock has remained stable in the last decade, sustaining a fishery of about 30000 t annually ( 15000 t in Greenland water). The East Greenland stock is a part of a stock complex extending from Greenland to the Barents

Sea. The stock size is currently estimated as being at a historical low. In 2015, catches were around 9400 t .

### 13.2.8 Lumpfish

The status of the lumpfish is unknown. The landing of lumpfish has increased dramatically in the last decades with catches being close to 13000 t in 2013. Catches are highest in the southern-mid section of the Greenland west coast. There are no indications of the impact on the stock. A management plan was implemented in 2014 regulating the fishery with TAC and number of fishing days.

### 13.2.9 Capelin

On the Greenland East coast an offshore pelagic fleet have been conducting a fishery on capelin (2500 t (summer/autumn) landed in 2015 by Greenland, EU, Norway and Iceland). The capelin has shifted distribution more west and north in recent years, and are believed to spend a substantial amount of time in Greenland waters. The west Greenland capelin stock is not fished and its size is unknown.

### 13.2.10 Mackerel

A mackerel fishery in Greenland waters initiated in 2011 with catches of 162 t and increased to more than 32000 t in 2015. Mackerel is known to feed on various species, including fish larvae, and it competes with others pelagic species, such as herring, for resources (Langøy et al. 2012). Thus it might/can have a key role on the ecosystem of many commercial important species in Greenland.

### 13.2.11 Herring

A fishery for Norwegian spring-spawning herring in Greenland water has increased in recent years and in 2014 catches increased to 9000 t . The herring has shifted distribution more west in recent years.

### 13.3 Advice on demersal fisheries

ICES recommends that the offshore cod stock is protected to allow for rebuilding. Inshore cod advice is based on the DLS approach. For the offshore cod, a recovery plan is recommended to ensure a sustainable increase in SSB and recruitment. Such initiatives must include appropriate measures to avoid any cod bycatch in other fisheries deploying mobile gears capable of catching cod. Observers must monitor functionalism of measures.

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## 14 Cod (Gadus morhua) in NAFO Subdivisions 1A-1E (Offshore West Greenland)

### 14.1 Stock definition

The cod found in Greenland is derived from four separate "stocks" that each is labelled by their spawning areas: I) offshore West Greenland waters; II) West Greenland fiords; III) offshore East Greenland and Icelandic waters and IV) inshore Icelandic waters (Therkildsen et al. 2013), (Figure 14.1).

From 2012 the inshore component (West Greenland, NAFO Subarea 1) was assessed separately from all offshore components. From 2015 the offshore West Greenland (NAFO subdivisions 1A-E) and East Greenland (NAFO subdivision 1F and ICES Subarea 14) components was assessed separately. The Stock Annex provides more details on the stock identities including the references to the primary literature.

### 14.2 Fishery

### 14.2.1 The emergence and collapse of the Greenland offshore cod fisheries

The Greenland commercial cod fishery in West Greenland started in the 1920s. The fishery gradually developed culminating with catch levels at 400000 tonnes annually in the 1960s. Due to overfishing and deteriorating environmental conditions the stock size declined and the fishery completely collapsed in the early 1990's (Tab 14.2.1, Fig 14.2.1). No fishery has developed since. More details on the historical development in the fisheries are provided in the stock annex.

### 14.2.2 The fishery in 2015

In 2014 a management plan for the offshore fisheries for cod was implemented with the overall objective of rebuilding the stock in West Greenland by closing the area for fisheries.

In 2015 the management plan was overruled and a TAC of 5000 tons was introduced as an experimental fishery.

Offshore catches in the fishery in 2015 amounted to a total of 4860 tons caught primarily on Dana Bank (NAFO div 1D and 1E, between $62^{\circ} 00-63^{\circ} 00 \mathrm{~N}$ ). Other areas of minor catch was Tovqussaq/Fyllas Bank north of Dana Bank (NAFO 1CD) and Narssalik Bank south of Dana Bank (NAFO 1E, figure 14.2.2.1 and 14.2.2.2).

The fishery occurred in spring August-December and peaked in November (table 14.2.2.1). Longliners took $36 \%$ of the total catches and trawlers took $64 \%$ of the total catch (figure 14.2.2.3).

### 14.2.3 Length, weight and age distributions in the fishery.

Length measurement amounted to 3555 cod measured. Length measurements were taken by crew members directly on the ships.

Overall mean length in the fishery was 62 cm and age 6 year old (YC 2009) dominated the catches followed by the 2010 YC (figure 14.2.3.1 and 14.2.3.2). The 2009 YC also dominated the catches in 2014 at age 5 years (table 14.2.3.1) and in 2015 dominated the catches in all areas. There were larger and older fish ( $7-8$ years) present on Dana Bank than on Tovqussaq bank furthest to the north in the fishing area in 2015 (table 7).

### 14.3 Surveys

At present, two offshore trawl surveys (Greenland and German) provide the core information relevant to stock assessment purposes. For details of survey design see stock annexes.

The German survey targets cod and has since 1982 covered the main cod grounds off West Greenland up to $67^{\circ} \mathrm{N}$ at depths down to 400 m , thus including periods of both high and low cod abundance. The Greenland survey targets shrimp and cod off West Greenland up to $72^{\circ} \mathrm{N}$ and from 0 to 600 m from 1992, hereby extending into northern areas where large cod concentrations are not expected. Although most of the effort has previously been allocated towards shrimp the recent addition of additional fish stations implies a fair coverage of the West Greenland cod habitat in this survey.

### 14.3.1 Results of the Greenland Shrimp and Fish survey

The numbers valid hauls were 185 in 2015 (table 14.3.1.1).
The 2015 survey abundance of Atlantic cod in West Greenland was estimated at 100 million individuals and the survey biomass at 95.487 tons (table 14.3.1.2 \& 14.3.1.3). Survey abundance decreased with $9 \%$ and biomass increased with $12 \%$ compared to 2014. Abundance was primarily in area 1C, D and E and biomass was primarily in NAFO Div. 1D and E (figure 14.3.1.1 and 14.3.1.2).

The stock was dominated by the 2009 YC in 2011, 2012 and 2013 accounting for $84 \%$, $64 \%$ and $52 \%$ of the total abundance respectively (table 14.3.1.4, figure 14,3,1,3). In 2014, the 2010 YC dominated the abundance with $51 \%$ of the total abundance followed by the 2009 YC accounting for $33 \%$ of the total abundance. In 2015, the 2010 YC was still the dominating YC accounting for $35 \%$ of the abundance followed by 2011 YC and 2009 YC that accounts for $21 \%$ each.

The 2009 YC was mainly found in the northern part of the survey (NAFO 1B) at age 2 in 2011 (figure 14.3.1.5). In 2012 and 2013, this YC was however mainly found in the southern part of the survey (NAFO Div. 1D and 1E). The 2010 and 2011 YC show the same distribution pattern of being in the northern part of the survey area (NAFO 1A and 1 B ) at ages 1 and 2 , and moving further to the south at ages 3 and 4 . Generally, younger YC's (age 1-2) are mainly found in the northern part of the survey area (NAFO Div. 1A and 1B).

The main cod found offshore in West Greenland are younger than 7 years, and the 2015 survey confirmed that older and larger cod barely exist offshore in West Greenland, however there is an increasing trend for cod at age 5-7. Especially the 6 years old fish ( 2009 YC) are very well represented compared to earlier years, representing the absolute highest abundance in the time-series (table 14.3.1.4, figure 14.3.1.3).

The offshore cod start to spawn at age 5-6 years, and the spawning-stock biomass in the survey show an increasing trend in recent years with spawning stock being concentrated in the southern area (NAFO 1E, figure 14.3.1.7 and 14.3.1.8). Recently, the number of spent females have indeed increased in the survey in area 1D and 1E (Figure 14.3.1.9).

The survey show a small decrease in abundance and an increase in biomass compared to 2014. The decrease in abundance might be caused by southward migration, of a fraction of the 2009 YC, to NAFO area 1F, as the survey in this area register increased numbers of this YC (Retzel 2016). The increase in biomass can be explained by the growth of the remaining 2009 YC and the following 2010 and 2011 YC.

### 14.3.2 Results of the German groundfish survey

In 2015, 35 valid trawl stations were sampled during autumn in the German Greenland offshore groundfish survey in West Greenland NAFO 1C-1E (Table 14.3.2.1).

Overall, abundance increased by 117\% from 2014-2015 (Table 14.3.2.2) and biomass increased by $154 \%$ (Table14.3.2.3). The main reason for the increase in abundance and biomass was one very large haul, located in NAFO 1D (figure 14.3.2.1 and 14.3.2.2) which contributed with $80 \%$ of the biomass estimate and resulted in high SD (table 14.3.2.3). Since 2012 the 2009 YC has dominated the survey and the 2010 has been the second most abundant YC in the survey. In 2015 the 2010 YC is the dominating YC at age 5 and the second largest YC is the 2009 and 2011 YC (age 6 and 4, table 14.3.2.4). These year classes are mainly observed in NAFO div. 1C and 1D in 2015 (figure 14.3.2.3) which is further to the north for the 2009 YC than in the Greenland survey.

The survey time-series shows three abundance peaks: one in 1987-1989 caused by the 1984 and 1985 YC, one in 2006 caused by the 2003 YC and one in 2012 caused by the 2009 YC (figure 14.3.2.4). Biomass indices show the same peaks, although an increase in biomass in the period 2012-2015 compared to the previous periods (figure 14.3.2.5).

Overall findings are the same in the Greenland and the German survey: the 2009 YC dominates the surveys in recent years and being record highest at age 6 in the timeseries. A 2010 YC that is dominating the survey in 2015 and the presence of a 2011 YC.

### 14.4 Information on spawning

No spawning of significance has been documented on the banks in West Greenland. In recent years', however, larger cod have been observed in the survey, especially in the southern part (NAFO 1E), and biomass is increasing. Especially the 2009 YC at age 6 is record with large numbers in the time-series of the 6 year old in 2015. Normally offshore cod start to spawn at age 6 , but whether spawning occurs in significant extent remains unknown since the survey is conducted outside spawning season. However, ogive state is noted in the survey and the number of spent females have increased in 2014 and 2015, indicating that some degree of spawning is occurring. Further investigation should be conducted to document the degree of spawning in the southern area, and to determine which stock these larger cod belong to.

### 14.5 Tagging experiments

A total of 17304 cod have been tagged in different regions of Greenland in the period of 2003-2015 (table 14.5.1). A total of 4604 cod in the offshore area in West Greenland have been tagged in 2007, 2012 and 2013 on Dana Bank (NAFO 1DE) and a small amount (57) was tagged further to the north on Tovqussaq bank (NAFO 1C) in 2015.

Offshore recaptures are found both in West- and East Greenland and Iceland (table 14.5.2). Tagged fish in the offshore area in West Greenland are more often caught in the same area ( 34 individuals), but some also migrate eastward (12 individuals recaptured in East Greenland, and 21 in Iceland, table 14.5.2). Limited fishing in several areas and years influences the signal from the recaptures, and more analysis needs to be performed taking the fishing effort into account in order to investigate magnitude of the eastward migration rate.

### 14.6 State of the stock

The West Greenland offshore stock component has been severely depleted since the 1970ies and collapsed in the 1990ies. The surveys show only a minor increase in biomass in recent years. Abundance however has fluctuated since 2005, indicating that small fish enter the survey but are not caught at older ages. This is caused by an eastward migration out of the area, and the area is currently considered to act mainly as a nursing area for the East Greenland and Icelandic stock components.

Recently the 2009 YC has been caught in considerable numbers and is believed to be of East Greenland and/or Icelandic origin and will probably migrate out of West Greenland when reaching maturity. However, at age 6 a part of this YC still remains in the southern part of the area (NAFO 1E), which has not been the typical pattern observed with the recent larger than average YC's from 2003, 2005 and 2007.

The stock is considered to be at a very low level compared to historic.

### 14.7 Implemented management measures for 2016

According to a management plan implemented in 2014 no offshore fishery is to take place in NAFO subdivision 1A-1E in 2016. The management plan has, however, been overruled, and a TAC of 5000 tons has been introduced.

### 14.8 Management plan

In 2014 a management plan was implemented for the offshore cod fishery in Greenland (2014-2016). The management plan is built on the distinction between the inshore and two offshore stocks components.


Management area West Greenland covers NAFO Subdivisions 1A-E and management area Southeast Greenland covers ICES Subarea 14.b (survey area Q1-6) + NAFO Subdivision 1 F corresponding to the ICES distinction.

According to the management plan, management area West TAC should be $0 t$ for the period 2014-2016 in order to protect the West offshore stock component. The TAC in management area Southeast is 10000 t /year between 2014 and 2016.
The management plan has not been evaluated by ICES.

### 14.9 Management considerations.

The fishery in West Greenland should be considered a mixed-stock fishery, containing fish from both Greenland and Iceland stocks. There is currently no standardized procedure to determine the proportional contribution of each stock to the landings. However, given the current state of the stock, catches taken in West Greenland waters will primarily consist of fish from other cod stocks.

The traditional spawning grounds in West Greenland are well described and if any fishing is allowed such areas should be protected. This will both protect any present spawning stock and minimize the proportion of the West Greenland stock in the catches.

### 14.10 Basis for advice

Basis for advice is the precautionary approach where biomass is extremely low and zero catch is advised.

### 14.11References

Retzel, A. 2016. Greenland Shrimp and Fish survey results for Atlantic cod in ICES Subarea 14.b (East Greenland) and NAFO subarea 1F (SouthWest Greenland) in 2015. ICES North Western Working Group (NWWG) April 27- May 4, 2016, WD 16.

Therkildsen, N.O.,Hemmer-Hansen, J., Hedeholm, R.B., Wisz, M.S., Pampoulie, C., Meldrup, D., Bonanomi, S., Retzel, A., Olsen, S.M., Nielsen, E.E. 2013. Spatiotemporal SNP analysis reveal pronounced biocomplexity at the northern renge margin of Atlantic cod Gadus morhua. Evoltutionary Applications. DOI 10.1111/eva. 12055

### 14.12Tables

Table 14.2.1. Offshore catches ( $\mathbf{t}$ ) divided into NAFO divisions in West Greenland. 1924-1991: Horsted 2000, 2004-present: Greenland Fisheries License Control.

| Year | NAFO <br> 1 A | NAFO 1 B | NAFO <br> 1 C | NAFO 1D | NAFO <br> 1E | NAFO <br> IF | Unknown NAFO DIV. | NAFO $1 A-1 E$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1924 |  |  |  |  |  |  | 200 |  |
| 1925 |  |  |  |  |  |  | 1871 |  |
| 1926 |  |  |  |  |  |  | 4452 |  |
| 1927 |  |  |  |  |  |  | 4427 |  |
| 1928 |  |  |  |  |  |  | 5871 |  |
| 1929 |  |  |  |  |  |  | 22304 |  |
| 1930 |  |  |  |  |  |  | 94722 |  |
| 1931 |  |  |  |  |  |  | 120858 |  |
| 1932 |  |  |  |  |  |  | 87273 |  |
| 1933 |  |  |  |  |  |  | 54351 |  |
| 1934 |  |  |  |  |  |  | 88422 |  |
| 1935 |  |  |  |  |  |  | 65796 |  |
| 1936 |  |  |  |  |  |  | 125972 |  |
| 1937 |  |  |  |  |  |  | 90296 |  |
| 1938 |  |  |  |  |  |  | 90042 |  |
| 1939 |  |  |  |  |  |  | 62807 |  |
| 1940 |  |  |  |  |  |  | 43122 |  |
| 1941 |  |  |  |  |  |  | 35000 |  |
| 1942 |  |  |  |  |  |  | 40814 |  |
| 1943 |  |  |  |  |  |  | 47400 |  |
| 1944 |  |  |  |  |  |  | 51627 |  |
| 1945 |  |  |  |  |  |  | 45800 |  |
| 1946 |  |  |  |  |  |  | 44395 |  |
| 1947 |  |  |  |  |  |  | 63458 |  |
| 1948 |  |  |  |  |  |  | 109058 |  |
| 1949 |  |  |  |  |  |  | 156015 |  |
| 1950 |  |  |  |  |  |  | 179398 |  |
| 1951 |  |  |  |  |  |  | 222340 |  |
| 1952 | 0 | 261 | 2996 | 18188 | 707 | 37905 | 257488 | 117126 * |
| 1953 | 4546 | 46546 | 10611 | 38915 | 932 | 25242 | 98225 | 180220 * |
| 1954 | 2811 | 97306 | 18192 | 91555 | 727 | 15350 | 60179 | 266682 * |
| 1955 | 773 | 50106 | 32829 | 87327 | 3753 | 4655 | 68488 | 241499 * |
| 1956 | 15 | 56011 | 38428 | 128255 | 8721 | 4922 | 66265 | 296315 * |
| 1957 | 0 | 58575 | 32594 | 62106 | 29093 | 16317 | 47357 | 225836* |
| 1958 | 168 | 55626 | 41074 | 73067 | 21624 | 26765 | 75795 | 258062 * |
| 1959 | 986 | 74304 | 10954 | 30254 | 12560 | 11009 | 67598 | 191343 * |
| 1960 | 35 | 58648 | 18493 | 35939 | 16396 | 9885 | 76431 | 200522 * |
| 1961 | 503 | 78018 | 43351 | 70881 | 16031 | 14618 | 90224 | 293104 * |
| 1962 | 1017 | 122388 | 75380 | 57972 | 25336 | 17289 | 125896 | 400719 * |
| 1963 | 66 | 70236 | 73142 | 76579 | 46370 | 16440 | 122653 | 381917 * |
| 1964 | 96 | 49049 | 49102 | 82936 | 33287 | 13844 | 99438 | 307878 * |
| 1965 | 385 | 80931 | 66817 | 71036 | 15594 | 15002 | 92630 | 321829 * |


| Year | NAFO 1A | NAFO <br> 1B | NAFO 1 C | NAFO <br> 1D | NAFO 1 E | NAFO <br> $1 F$ | Unknown NAFO DIV. | $\begin{aligned} & \text { NAFO } \\ & 1 A-1 E \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 12 | 99495 | 43557 | 62594 | 19579 | 18769 | 95124 | 313044 * |
| 1967 | 361 | 58612 | 78270 | 122518 | 34096 | 12187 | 95911 | 385949 * |
| 1968 | 881 | 12333 | 89636 | 94820 | 61591 | 16362 | 97390 | 350870 * |
| 1969 | 490 | 7652 | 31140 | 65115 | 41648 | 11507 | 35611 | 179055 * |
| 1970 | 278 | 3719 | 13244 | 23496 | 23215 | 15519 | 18420 | 78775 * |
| 1971 | 39 | 1621 | 28839 | 21188 | 9088 | 20515 | 26384 | 80501* |
| 1972 | 0 | 3033 | 42736 | 18699 | 7022 | 4396 | 20083 | 90410* |
| 1973 | 0 | 2341 | 17735 | 18587 | 10581 | 2908 | 1168 | 50347 * |
| 1974 | 36 | 1430 | 12452 | 14747 | 8701 | 1374 | 656 | 37999 * |
| 1975 | 0 | 49 | 18258 | 12494 | 6880 | 3124 | 549 | 38188* |
| 1976 | 0 | 442 | 5418 | 10704 | 8446 | 2873 | 229 | 25215 * |
| 1977 | 127 | 301 | 4472 | 7943 | 8506 | 2175 | $35477{ }^{1}$ | 53546 * |
| 1978 | 0 | 0 | 11856 | 2638 | 3715 | 549 | $34563{ }^{1}$ | 51760 * |
| 1979 | 0 | 16 | 6561 | 4042 | 1115 | 537 | $51139{ }^{1}$ | 60635 * |
| 1980 | 0 | 1800 | 2200 | 2117 | 1687 | 384 | $7241{ }^{1}$ | 14705 * |
| 1981 | 0 | 0 | 4289 | 4701 | 4508 | 255 | 0 | 13498 |
| 1982 | 0 | 133 | 6143 | 10977 | 11222 | 692 | 1174 | 29621 * |
| 1983 | 0 | 0 | 717 | 6223 | 16518 | 4628 | 293 | 23703 * |
| 1984 | 0 | 0 | 0 | 4921 | 5453 | 3083 | 0 | 10374 |
| 1985 | 0 | 0 | 0 | 145 | 1961 | 1927 | 2402 | 3360 * |
| 1986 | 0 | 0 | 0 | 2 | 72 | 24 | 1203 | 982 * |
| 1987 | 0 | 0 | 5 | 815 | 67 | 43 | 3041 | 3787* |
| 1988 | 0 | 0 | 919 | 17463 | 10913 | 6466 | 8101 | 35931 * |
| 1989 | 0 | 0 | 0 | 11071 | 48092 | 14248 | 2 | 59165 |
| 1990 | 0 | 0 | 2 | 563 | 21513 | 10580 | 7503 | 27151 * |
| 1991 | 0 | 0 | 0 | 0 | 104 | 1942 | 0 | 104 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 5 | 3 | 1 | 0 | 8 |
| 2005 | 0 | 0 | 1 | 0 | 0 | 71 | 0 | 1 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 414 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 31 | 435 | $2011{ }^{2}$ | 0 | 466 |
| 2008 | 0 | 0 | 0 | 23 | 526 | $11370{ }^{2}$ | 0 | 549 |
| 2009 | 0 | 0 | 0 | 0 | 6 | $3323{ }^{2}$ | 0 | 6 |
| 2010 | 0 | 0 | 0 | 0 | 2 | 281 | 0 | 2 |
| 2011 | 0 | 0 | 0 | 0 | 8 | 542 | 0 | 8 |


| Year | NAFO <br> 1 A | NAFO 1 B | NAFO <br> 1 C | NAFO <br> 1D | NAFO 1 E | NAFO $1 F$ | Unknown NAFO DIV. | NAFO $1 A-1 E$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 0 | 0 | 1 | 95 | 236 | 1470 | 0 | 332 |
| 2013 | 0 | 0 | 0 | 209 | 270 | 1405 | 0 | 479 |
| 2014 | 0 | 0 | 30 | 68 | 18 | 1833 | 0 | 116 |
| 2015 | 0 | 0 | 341 | 954 | 3564 | 3984 | 0 | 4860 |

Estimates for assessment include estimates of unreported catches. The total estimated value for West Greenland (inshore + offshore) was 73000 t in 1977 and 1978, 1979: 99000 t, 1980: 54000 t . The value given in the table are these values minus the inshore catches minus known offshore NAFO division catches.

Include catches taken with small vessels and landed to a factory in South Greenland (Qaqortoq), 2007:597 t, 2008: 2262 t, 2009: 136 t.
*Unknown NAFO division catches added accordingly to the proportion of known catch in NAFO divisions 1A-1E to known total catch in all NAFO divisions.

Table 14.2.2.1: 2015 cod catches ( $\mathbf{t}$ ) divided into month and NAFO areas, caught by the offshore fisheries.

| NAFO | JAN | Feb | Mar | APR | May | Jun | JuL | Aug | Sep | Ост | Nov | Dec | Total | \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1C |  |  |  |  |  |  |  |  |  | 1 | 163 | 177 | 341 | 7\% |
| 1D |  |  |  |  |  |  |  | 28 | 4 | 305 | 615 | 2 | 954 | 20\% |
| 1E |  |  |  |  |  |  | 4 | 302 | 746 | 878 | 1016 | 618 | 3564 | 73\% |
| Total |  |  |  |  |  |  | 4 | 330 | 749 | 1184 | 1795 | 798 | 4860 |  |
| \% |  |  |  |  |  |  | 0.1\% | 7\% | 15\% | 24\% | 37\% | 16\% |  |  |

Table 14.2.2.3: 2015 cod catches ( $t$ ) by gear, area and month in Westgreenland.

| Gear | NAFO | JAN | Feb | Mar | APR | MAY | Jun | JUL | Aug | Sep | Oct | Nov | Dec | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Longline | 1D |  |  |  |  |  |  |  | 28 |  | 304 | 242 |  | 574 |
|  | 1E |  |  |  |  |  |  |  | 46 | 437 | 265 | 297 | 127 | 1172 |
|  | Total |  |  |  |  |  |  |  | 74 | 437 | 570 | 539 | 127 | 1746 |
| Trawl | 1C |  |  |  |  |  |  |  |  |  | 1 | 163 | 177 | 341 |
|  | 1D |  |  |  |  |  |  |  |  | 4 | 1 | 373 | 2 | 379 |
|  | 1E |  |  |  |  |  |  | 4 | 256 | 309 | 613 | 720 | 491 | 2392 |
|  | Total |  |  |  |  |  |  | 4 | 256 | 312 | 614 | 1256 | 671 | 3113 |

Table 14.2.3.1. Cod in Greenland. Catch-at-age ('000) and Weight at age (kg) for offshore fleets in Westgreenland (NAFO 1A-1E). Yellow highlights dominating year classes in the catches.

| CATCH-AT-AGE |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year/age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 2007 | 6 | 167 | 66 | 42 | 6 | 1 |  |  |
| $2008$ |  |  |  |  |  |  |  |  |
| $2009$ |  |  |  |  |  |  |  |  |
| 2010 |  |  |  |  |  |  |  |  |
| 2011 |  |  |  |  |  |  |  |  |
| 2012 | 8 | 33 | 107 | 38 | 18 | 2 | 0.01 | 0.003 |
| 2013 |  | 15 | 44 | 113 | 29 | 15 | 4 | 1 |
| 2014 | 1 | 18 | 45 | 7 | 9 | 2 | 0.02 |  |
| 2015 | 6 | 67 | 502 | 1061 | 240 | 158 | 45 | 16 |
| Weight at age |  |  |  |  |  |  |  |  |
| 2007 | 0.647 | 0.906 | 1.949 | 3.440 | 5.817 | 6.053 |  |  |
| $2008$ |  |  |  |  |  |  |  |  |
| $2009$ |  |  |  |  |  |  |  |  |
| 2010 |  |  |  |  |  |  |  |  |
| $2011$ |  |  |  |  |  |  |  |  |
| 2012 | 0.560 | 0.935 | 1.395 | 2.139 | 3.232 | 4.194 | 8.325 | 12.500 |
| 2013 |  | 1.120 | 1.462 | 1.947 | 2.978 | 3.754 | 6.398 | 7.342 |
| 2014 | 0.488 | 0.693 | 1.199 | 1.738 | 3.040 | 4.817 | 5.318 |  |
| 2015 | 0.474 | 0.734 | 1.316 | 1.982 | 3.186 | 5.043 | 7.167 | 10.329 |

Table 14.3.1.1. Number of hauls in the Greenland Shrimp and Fish survey in West Greenland by year and NAFO subdivisions.

| West Greenland |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year/NAFO | 0A | 1 A | 1 B | 1 C | 1 D | 1E | Total |
| 1992 |  | 92 | 44 | 18 | 18 | 11 | 183 |
| 1993 |  | 69 | 49 | 21 | 15 | 12 | 166 |
| 1994 |  | 76 | 58 | 23 | 8 | 9 | 174 |
| 1995 |  | 83 | 61 | 29 | 13 | 14 | 200 |
| 1996 |  | 71 | 57 | 29 | 12 | 9 | 178 |
| 1997 |  | 84 | 56 | 32 | 12 | 12 | 196 |
| 1998 |  | 77 | 80 | 27 | 19 | 14 | 217 |
| 1999 |  | 84 | 81 | 33 | 16 | 14 | 228 |
| 2000 |  | 56 | 62 | 37 | 23 | 14 | 192 |
| 2001 |  | 60 | 75 | 36 | 24 | 15 | 210 |
| 2002 |  | 50 | 80 | 32 | 18 | 20 | 200 |
| 2003 |  | 51 | 63 | 30 | 18 | 15 | 177 |
| 2004 |  | 54 | 55 | 24 | 22 | 20 | 175 |
| New Survey Gear Introduced |  |  |  |  |  |  |  |
| 2005 | 6 | 65 | 56 | 26 | 19 | 23 | 195 |
| 2006 | 5 | 86 | 60 | 26 | 20 | 21 | 218 |
| 2007 | 8 | 73 | 58 | 26 | 27 | 31 | 223 |
| 2008 | 6 | 69 | 61 | 28 | 23 | 25 | 212 |
| 2009 | 8 | 74 | 75 | 28 | 22 | 24 | 231 |
| 2010 | 10 | 95 | 76 | 30 | 23 | 25 | 259 |
| 2011 | 0 | 73 | 64 | 24 | 18 | 12 | 191 |
| 2012 | 0 | 73 | 64 | 21 | 18 | 18 | 194 |
| 2013 | 4 | 73 | 52 | 20 | 13 | 21 | 183 |
| 2014 | 0 | 78 | 57 | 19 | 17 | 23 | 194 |
| 2015 | 0 | 70 | 49 | 24 | 22 | 20 | 185 |

Table 14.3.1.2 Cod abundance indices ('000) from the Greenland Shrimp and Fish survey in West Greenland by year and NAFO subdivisions.

| West Greentand |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | OA | 1 A | 1 B | 1 C | 1D | 1E | Total | CV |
| 1992 |  | 4 | 53 | 243 | 345 | 0 | 645 |  |
| 1993 |  | 2 | 16 | 54 | 135 | 286 | 493 |  |
| 1994 |  | 10 | 41 | 87 | 0 | 6 | 144 |  |
| 1995 |  | 0 | 51 | 380 | 44 | 62 | 537 |  |
| 1996 |  | 0 | 0 | 46 | 68 | 87 | 201 |  |
| 1997 |  | 0 | 7 | 31 | 0 | 0 | 38 |  |
| 1998 |  | 0 | 4 | 0 | 26 | 26 | 56 |  |
| 1999 |  | 32 | 136 | 16 | 23 | 6 | 213 |  |
| 2000 |  | 585 | 437 | 71 | 58 | 9 | 1160 |  |
| 2001 |  | 26 | 305 | 110 | 448 | 305 | 1194 |  |
| 2002 |  | 13 | 203 | 78 | 3294 | 114 | 3702 |  |
| 2003 |  | 492 | 1395 | 351 | 727 | 214 | 3179 |  |
| 2004 |  | 197 | 152 | 379 | 2630 | 1538 | 4896 |  |
| New Survey Gear Introduced |  |  |  |  |  |  |  |  |
| 2005 | 143 | 198 | 871 | 1845 | 4796 | 6683 | 14537 | 25 |
| 2006 | 453 | 371 | 4454 | 2564 | 15703 | 3359 | 26905 | 45 |
| 2007 | 737 | 1318 | 3302 | 7353 | 3624 | 3296 | 19628 | 31 |
| 2008 | 1209 | 897 | 4185 | 4068 | 9008 | 11553 | 30913 | 27 |
| 2009 | 881 | 889 | 4195 | 3272 | 2788 | 1252 | 13277 | 12 |
| 2010 | 338 | 720 | 2837 | 2712 | 8295 | 2745 | 17647 | 23 |
| 2011 |  | 8756 | 47092 | 2179 | 26510 | 1013 | 85549 | 14 |
| 2012 |  | 7661 | 10228 | 3017 | 1270 | 27081 | 49258 | 54 |
| 2013 | 4613 | 8951 | 12864 | 5673 | 7887 | 29924 | 69911 | 43 |
| 2014 |  | 6911 | 5670 | 78854 | 2456 | 16254 | 110145 | 67 |
| 2015 |  | 6542 | 11213 | 27248 | 31703 | 23493 | 100198 | 34 |

Table 14.3.1.3. Cod biomass indices (tons) from the Greenland Shrimp and Fish survey in West Greenland by year and NAFO subdivisions.

| West Greenland |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | OA | 1A | 1B | 1 C | 1 D | 1E | Total | CV |
| 1992 |  | 23 | 54 | 75 | 118 | 0 | 270 |  |
| 1993 |  | 2 | 5 | 25 | 39 | 124 | 195 |  |
| 1994 |  | 3 | 9 | 38 | 0 | 1 | 51 |  |
| 1995 |  | 5 | 6 | 120 | 23 | 3 | 157 |  |
| 1996 |  | 0 | 0 | 15 | 23 | 27 | 65 |  |
| 1997 |  | 0 | 2 | 53 | 0 | 0 | 55 |  |
| 1998 |  | 1 | 1 | 0 | 47 | 50 | 99 |  |
| 1999 |  | 29 | 28 | 1 | 17 | 1 | 76 |  |
| 2000 |  | 226 | 130 | 21 | 9 | 2 | 388 |  |
| 2001 |  | 140 | 155 | 56 | 178 | 98 | 627 |  |
| 2002 |  | 67 | 128 | 41 | 1489 | 42 | 1767 |  |
| 2003 |  | 444 | 323 | 264 | 453 | 118 | 1602 |  |
| 2004 |  | 542 | 53 | 176 | 680 | 685 | 2136 |  |
| New Surver Gear Introduced |  |  |  |  |  |  |  |  |
| 2005 | 38 | 69 | 364 | 458 | 1084 | 1141 | 3155 | 26 |
| 2006 | 114 | 62 | 677 | 537 | 5131 | 525 | 7046 | 64 |
| 2007 | 247 | 387 | 872 | 1562 | 628 | 659 | 4355 | 31 |
| 2008 | 413 | 377 | 2046 | 929 | 1633 | 3227 | 8625 | 28 |
| 2009 | 208 | 230 | 1251 | 711 | 439 | 253 | 3092 | 14 |
| 2010 | 180 | 263 | 999 | 543 | 2426 | 908 | 5319 | 22 |
| 2011 |  | 1569 | 9654 | 408 | 5316 | 191 | 17140 | 14 |
| 2012 |  | 1932 | 2938 | 1125 | 464 | 14103 | 20562 | 69 |
| 2013 | 2395 | 2692 | 3960 | 1732 | 4551 | 19017 | 34345 | 53 |
| 2014 |  | 2639 | 2305 | 56061 | 2511 | 21381 | 84897 | 64 |
| 2015 |  | 3463 | 4456 | 19705 | 33169 | 34695 | 95487 | 38 |

Table 14.3.1.4: Abundance indices (' ${ }^{\prime} 000$ ) by year class/age from the Greenland Shrimp and Fish survey in West Greenland (NAFO 1A-1E).

| WEST GREENLAND |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR/AGE | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| 2005 | 134 | 815 | 10247 | 1604 | 1514 | 186 | 35 | 2 | 0 | 0 | 0 |
| 2006 | 249 | 6543 | 3577 | 12677 | 3395 | 401 | 47 | 16 | 0 | 0 | 0 |
| 2007 | 152 | 270 | 13792 | 3439 | 1934 | 37 | 4 | 0 | 0 | 0 | 0 |
| 2008 | 31 | 3472 | 2692 | 18780 | 4904 | 868 | 121 | 44 | 0 | 0 | 0 |
| 2009 | 0 | 124 | 9442 | 1666 | 1717 | 326 | 3 | 0 | 0 | 0 | 0 |
| 2010 | 209 | 2703 | 2094 | 10566 | 1252 | 775 | 42 | 7 | 0 | 0 | 0 |
| 2011 | 19 | 4940 | 71837 | 4453 | 3735 | 391 | 175 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 204 | 11264 | 31593 | 3648 | 2427 | 116 | 7 | 0 | 0 | 0 |
| 2013 | 0 | 2904 | 8912 | 15168 | 36226 | 5665 | 848 | 142 | 22 | 25 | 0 |
| 2014 | 0 | 471 | 4792 | 8088 | 56469 | 35839 | 2597 | 1718 | 125 | 35 | 11 |
| 2015 | 0 | 2210 | 3932 | 15038 | 21509 | 34766 | 21117 | 1196 | 348 | 70 | 12 |

Table 14.3.1.5 Abundance indices ('000) by age from the Greenland Shrimp and Fish survey in West Greenland by NAFO divisions, 2015.

| West Greenland |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year class | 2015201420132 |  |  | 20122011 |  | 2010 | 2009 |  | 2007 | 2006<2006 |  |
| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| Div. 0A |  |  |  |  |  |  |  |  |  |  |  |
| Div. 1A | 0 | 149 | 1037 | 3066 | 1107 | 939 | 175 | 69 | 0 | 0 | 0 |
| Div. 1B | 0 | 1758 | 2245 | 3042 | 2469 | 1251 | 436 | 12 | 0 | 0 | 0 |
| Div. 1C | 0 | 205 | 569 | 6265 | 8255 | 8337 | 3404 | 184 | 29 | 0 | 0 |
| Div. 1D | 0 | 98 | 41 | 2000 | 7177 | 16214 | 5940 | 185 | 0 | 48 | 0 |
| Div. 1E | 0 | 0 | 39 | 665 | 2502 | 8026 | 11161 | 746 | 319 | 22 | 12 |

Table 14.3.2.1 German survey. Numbers of valid hauls by stratum in West Greenland (NAFO 1C-E).

|  | NAFO 1 C |  | NAFO 1D |  | NAFO 1E |  | Sum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | STR 1.1 | STR 1.2 | STR 2.1 | STR 2.2 | STR 3.1 | STR 3.2 |  |
| 1981 | 1 | 1 | 13 | 2 | 3 | 1 | 21 |
| 1982 | 20 | 11 | 16 | 7 | 9 | 6 | 69 |
| 1983 | 26 | 11 | 25 | 11 | 17 | 5 | 95 |
| 1984 | 25 | 13 | 26 | 8 | 19 | 6 | 97 |
| 1985 | 10 | 8 | 26 | 10 | 17 | 5 | 76 |
| 1986 | 27 | 9 | 21 | 9 | 16 | 7 | 89 |
| 1987 | 25 | 19 | 21 | 4 | 18 | 4 | 91 |
| 1988 | 34 | 21 | 28 | 5 | 18 | 5 | 111 |
| 1989 | 25 | 14 | 30 | 9 | 8 | 3 | 89 |
| 1990 | 19 | 7 | 23 | 8 | 16 | 3 | 76 |
| 1991 | 19 | 11 | 23 | 7 | 13 | 6 | 79 |
| 1992 | 6 | 6 | 6 | 5 | 6 | 6 | 35 |
| 1993 | 9 | 7 | 9 | 6 | 10 | 8 | 49 |
| 1994 | 16 | 13 | 13 | 8 | 10 | 6 | 66 |
| 1995 | . | . | 3 | . | 10 | 7 | 20 |
| 1996 | 5 | 5 | 8 | 5 | 12 | 5 | 40 |
| 1997 | 5 | 6 | 5 | 5 | 6 | 5 | 32 |
| 1998 | 9 | 5 | 10 | 7 | 11 | 6 | 48 |
| 1999 | 8 | 7 | 14 | 8 | 13 | 6 | 56 |
| 2000 | 13 | 6 | 15 | 6 | 14 | 5 | 59 |
| 2001 | . | . | 15 | 7 | 15 | 5 | 42 |
| 2002 | . | . | 7 | 2 | 5 | 6 | 20 |
| 2003 | . | . | 7 | 6 | 7 | 7 | 27 |
| 2004 | 8 | 8 | 11 | 9 | 9 | 5 | 50 |
| 2005 | . | . | 9 | 7 | 8 | 6 | 30 |
| 2006 | 6 | 5 | 7 | 5 | 7 | 7 | 37 |
| 2007 | 5 | 5 | 7 | 5 | 6 | 5 | 33 |
| 2008 | 5 | . | 7 | 7 | 7 | 9 | 35 |
| 2009 | 2 | . | 5 | 5 | 6 | 6 | 24 |
| 2010 | 5 | 5 | 10 | 5 | 7 | 9 | 41 |
| 2011 | . | . | 5 | 5 | 5 | 5 | 20 |
| 2012 | 5 | 5 | 10 | 8 | 9 | 7 | 44 |
| 2013 | 6 | 6 | 8 | 6 | 10 | 7 | 43 |
| 2014 | 5 | 5 | 10 | 8 | 10 | 7 | 45 |
| 2015 | 7 | 7 | 7 | 4 | 5 | 5 | 35 |

Table 14.3.2.2 German survey. Cod abundance indices ('000) from the German survey in West Greenland (NAFO 1C-1E) by year and stratum.

| NAFO 1C |  |  |  | NAFO 1D |  | NAFO 1E |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| year | str1_1 | str1_2 | str2_1 | str2_2 | str3_1 | str3_2 | Sum | SD |
| 1982 | 2364 | 408 | 27594 | 920 | 7401 | 1801 | 40488 | 18605 |
| 1983 | 177 | 196 | 7079 | 2230 | 8678 | 1230 | 19590 | 7266 |


| 1984 | 189 | 90 | 2524 | 98 | 2666 | 364 | 5931 | 3629 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1985 | 8094 | 1107 | 7237 | 2348 | 4984 | 840 | 24610 | 10809 |
| 1986 | 14716 | 630 | 22985 | 108 | 16570 | 609 | 55618 | 29631 |
| 1987 | 173517 | 482 | 115172 | 3790 | 72349 | 186 | 365496 | 331763 |
| 1988 | 46027 | 1106 | 186523 | 43090 | 21037 | 51 | 297834 | 216925 |
| 1989 | 1362 | 483 | 16280 | 325 | 129005 | 678 | 148133 | 65933 |
| 1990 | 619 | 299 | 2279 | 235 | 3827 | 61 | 7320 | 5462 |
| 1991 | 142 | 116 | 88 | 92 | 474 | 387 | 1299 | 412 |
| 1992 | 274 | 334 | 72 | 127 | 57 | 38 | 902 | 314 |
| 1993 | 327 | 243 | 105 | 109 | 53 | 21 | 858 | 195 |
| 1994 | 95 | 53 | 16 | 17 | 34 | 11 | 226 | 79 |
| 1995 | $\cdot$ | $\cdot$ | 27 | $\cdot$ | 72 | 34 | 133 | 60 |
| 1996 | 82 | 70 | 42 | 20 | 65 | 0 | 279 | 80 |
| 1997 | 0 | 24 | 17 | 0 | 57 | 3 | 101 | 45 |
| 1998 | 793 | 0 | 23 | 28 | 7 | 0 | 851 | 573 |
| 1999 | 103 | 33 | 33 | 11 | 197 | 7 | 384 | 171 |
| 2000 | 205 | 250 | 50 | 174 | 288 | 9 | 976 | 383 |
| 2001 | . | $\cdot$ | 584 | 36 | 3020 | 9 | 3649 | 3481 |
| 2002 | . | . | 238 | 21 | 342 | 23 | 624 | 257 |
| 2003 | $\cdot$ | . | 625 | 99 | 1625 | 73 | 2422 | 945 |
| 2004 | 503 | 213 | 1522 | 123 | 2709 | 638 | 5708 | 1592 |
| 2005 | $\cdot$ | $\cdot$ | 1586 | 264 | 5666 | 419 | 7935 | 3115 |
| 2006 | 495 | 485 | 87439 | 858 | 4481 | 1323 | 95081 | 99523 |
| 2007 | 1430 | 3261 | 3417 | 687 | 9861 | 71 | 18727 | 8645 |
| 2008 | 2666 | $\cdot$ | 916 | 911 | 23527 | 616 | 28636 | 26712 |
| 2009 | 72 | $\cdot$ | 1370 | 850 | 1068 | 378 | 3738 | 879 |
| 2010 | 2644 | 464 | 4451 | 631 | 5148 | 274 | 13612 | 6231 |
| 2011 | . | $\cdot$ | 716 | 375 | 1242 | 337 | 2670 | 782 |
| 2012 | 99609 | 1253 | 6007 | 442 | 8455 | 1251 | 117017 | 68441 |
| 2013 | 4457 | 1585 | 20122 | 221 | 7138 | 252 | 33775 | 22438 |
| 2014 | 9952 | 2008 | 28102 | 413 | 1261 | 86 | 41822 | 38616 |
| 2015 | 13315 | 906 | 73434 | 471 | 2432 | 102 | 90660 | 73453 |
|  |  |  |  |  |  |  |  |  |

Table 14.3.2.3 German survey, Cod biomass indices (tons) from the German survey in West Greenland (NAFO 1C-1E) by year and stratum.

| NAFO 1C |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| year | str1_1 | str1_2 | str2_1 | str2_2 | str3_1 | str3_2 | Sum | SD |
| 1982 | 1113 | 163 | 37404 | 1280 | 9970 | 4483 | 54413 | 26014 |
| 1983 | 144 | 87 | 9052 | 3381 | 12953 | 5015 | 30632 | 10295 |
| 1984 | 406 | 104 | 3998 | 137 | 3643 | 551 | 8839 | 5507 |
| 1985 | 1046 | 112 | 6543 | 1181 | 4700 | 506 | 14088 | 18209 |
| 1986 | 4858 | 254 | 11787 | 36 | 12381 | 651 | 29967 | 13885 |
| 1987 | 148896 | 156 | 93292 | 2446 | 54178 | 107 | 299075 | 299459 |
| 1988 | 47085 | 579 | 190073 | 39548 | 19663 | 54 | 297002 | 227428 |


| 1989 | 384 | 124 | 15061 | 211 | 113614 | 710 | 130104 | 55334 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1990 | 130 | 66 | 1948 | 123 | 3652 | 56 | 5975 | 4986 |
| 1991 | 45 | 38 | 36 | 28 | 549 | 374 | 1070 | 529 |
| 1992 | 65 | 104 | 15 | 33 | 10 | 7 | 234 | 97 |
| 1993 | 77 | 45 | 27 | 27 | 30 | 6 | 212 | 53 |
| 1994 | 13 | 17 | 3 | 12 | 11 | 5 | 61 | 17 |
| 1995 | $\cdot$ | $\cdot$ | 14 | $\cdot$ | 13 | 7 | 34 | 12 |
| 1996 | 13 | 35 | 12 | 11 | 28 | 0 | 99 | 29 |
| 1997 | 0 | 21 | 11 | 0 | 50 | 3 | 85 | 43 |
| 1998 | 38 | 0 | 1 | 7 | 1 | 0 | 47 | 25 |
| 1999 | 16 | 11 | 6 | 3 | 63 | 5 | 104 | 57 |
| 2000 | 54 | 71 | 11 | 83 | 73 | 5 | 297 | 117 |
| 2001 | $\cdot$ | $\cdot$ | 163 | 17 | 1024 | 5 | 1209 | 1212 |
| 2002 | $\cdot$ | $\cdot$ | 89 | 16 | 136 | 7 | 248 | 108 |
| 2003 | $\cdot$ | $\cdot$ | 98 | 44 | 736 | 32 | 910 | 461 |
| 2004 | 172 | 83 | 274 | 45 | 547 | 186 | 1307 | 342 |
| 2005 | $\cdot$ | $\cdot$ | 605 | 124 | 1796 | 146 | 2671 | 1057 |
| 2006 | 102 | 138 | 45616 | 250 | 2046 | 614 | 48766 | 52298 |
| 2007 | 319 | 885 | 1579 | 244 | 7804 | 43 | 10874 | 7524 |
| 2008 | 872 | $\cdot$ | 193 | 206 | 11479 | 175 | 12925 | 13686 |
| 2009 | 19 | $\cdot$ | 309 | 293 | 372 | 153 | 1146 | 255 |
| 2010 | 1012 | 244 | 2234 | 312 | 2703 | 173 | 6678 | 3057 |
| 2011 | $\cdot$ | $\cdot$ | 189 | 128 | 1040 | 194 | 1551 | 602 |
| 2012 | 52497 | 588 | 4185 | 240 | 8203 | 848 | 66561 | 35693 |
| 2013 | 2703 | 1670 | 17316 | 142 | 11251 | 544 | 33626 | 18801 |
| 2014 | 10597 | 2154 | 35741 | 422 | 3561 | 397 | 52872 | 47451 |
| 2015 | 17221 | 1105 | 109073 | 522 | 5999 | 216 | 134136 | 108717 |
|  |  |  |  |  |  |  |  |  |

Table 14.3.2.4 German survey, West Greenland (NAFO 1C-1D). Age disaggregated abundance indices ('1000).

| Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1 +}$ | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 |  | 77 | 505 | 14266 | 5195 | 14798 | 4144 | 908 | 178 | 344 | 35 | 34 | 40484 |
| $\left.1983^{*}\right)$ |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1984 | 80 | 3 | 13 | 709 | 604 | 3495 | 289 | 628 | 32 | 61 | 13 | 0 | 5927 |
| 1985 | 202 | 16823 | 623 | 330 | 2271 | 1100 | 2982 | 112 | 164 | 2 | 3 | 0 | 24612 |
| 1986 |  | 3600 | 45772 | 1686 | 321 | 2386 | 652 | 1098 | 22 | 74 | 3 | 1 | 55615 |
| 1987 |  | 147 | 22578 | 318948 | 13977 | 2930 | 4603 | 649 | 1506 |  | 131 | 13 | 365482 |
| 1988 |  | 124 | 1357 | 44364 | 247618 | 2660 | 311 | 521 | 318 | 529 | 12 | 15 | 297829 |
| 1989 | 0 | 163 | 1293 | 3821 | 79642 | 62126 | 1008 |  | 47 | 7 | 24 | 0 | 148131 |
| 1990 | 11 | 17 | 595 | 1242 | 368 | 4089 | 990 | 6 | 0 | 0 |  | 1 | 7319 |
| 1991 |  | 86 | 94 | 193 | 350 | 36 | 461 | 57 | 2 |  |  | 0 | 1279 |
| 1992 | 88 | 672 | 100 | 17 | 25 |  | 0 |  |  |  | 0 | 902 |  |
| 1993 | 8 | 499 | 318 | 12 | 21 |  |  |  |  |  | 0 | 858 |  |
| 1994 | 98 | 18 | 90 | 14 | 3 |  | 2 |  |  |  | 0 | 225 |  |
| 1995 |  | 111 | 6 | 16 |  |  |  |  |  |  | 0 | 133 |  |


| 1996 |  | 76 | 6 | 193 | 5 |  | 0 |  |  |  |  | 0 | 280 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 |  | 6 | 13 | 7 | 76 |  |  |  |  |  |  | 0 | 102 |
| 1998 | 0 | 845 |  | 3 | 3 | 0 |  |  |  |  |  | 0 | 851 |
| 1999 | 8 | 165 | 166 | 36 | 3 |  | 3 |  |  |  |  | 0 | 381 |
| 2000 |  | 60 | 524 | 328 | 62 |  |  |  |  |  |  | 0 | 974 |
| 2001 |  | 266 | 2753 | 527 | 65 | 20 |  |  |  |  |  | 0 | 3631 |
| 2002 | 0 | 6 | 309 | 290 | 17 |  |  |  |  |  |  | 0 | 622 |
| 2003 |  | 1368 | 205 | 511 | 284 | 36 | 9 |  |  |  |  | 0 | 2413 |
| 2004 | 132 | 3078 | 2008 | 307 | 108 | 55 | 15 | 0 |  |  |  | 0 | 5703 |
| 2005 | 91 | 156 | 6893 | 653 | 40 | 16 | 14 | 0 | 0 |  |  | 0 | 7863 |
| 2006 | 157 | 1949 | 6961 | 83106 | 2708 | 45 | 51 | 67 | 0 |  |  | 0 | 95044 |
| 2007 | 139 | 229 | 9402 | 1655 | 6989 | 227 | 35 | 38 | 12 |  |  | 0 | 18726 |
| 2008 | 8 | 1224 | 2317 | 20080 | 3747 | 1235 | 20 | 3 | 2 | 0 | 0 | 0 | 28636 |
| 2009 | 36 | 326 | 2513 | 363 | 406 | 37 | 40 | 14 |  |  |  | 0 | 3735 |
| 2010 | 208 | 1531 | 1726 | 9201 | 577 | 259 | 51 | 48 | 3 | 3 |  | 5 | 13612 |
| 2011 |  | 195 | 1572 | 385 | 368 | 68 | 33 | 26 | 24 | 0 | 0 | 0 | 2671 |
| 2012 | 142 | 1191 | 37872 | 66947 | 7682 | 2847 | 227 | 76 | 8 | 18 |  | 0 | 117010 |
| 2013 |  | 152 | 1562 | 12824 | 15859 | 1783 | 1135 | 234 | 86 | 23 | 18 | 4 | 33680 |
| 2014 |  |  | 880 | 4629 | 17021 | 17863 | 1080 | 277 | 32 | 0 | 4 | 0 | 41786 |
| 2015 | 159 | 189 | 1353 | 10921 | 16208 | 43991 | 16909 | 708 | 87 | 117 | 8 | 12 | 90660 |

*) calculated proportionally using age compositions reported by the ICES Working Group on Cod Stocks off East Greenland (ICES, 1984).

Table 14.5.1. Number of tagged cod in the period of 2003 to 2015 in different regions. Bank (West) = NAFO division 1D+1E. East Greenland = NAFO division 1F + ICES division XIVb.

| TAGGED |  |  |  |
| :---: | :---: | :---: | :---: |
| Year | FJord | Bank (West) | EAST Greenland |
| 2003 | 599 |  |  |
| 2004 | 658 |  |  |
| 2005 | 565 |  | 1387 |
| 2006 | 41 | 721 | 1296 |
| 2007 | 1140 |  | 525 |
| 2008 | 231 |  | 403 |
| 2009 | 633 |  | 2359 |
| 2010 | 88 |  |  |
| 2011 | 28 |  |  |
| 2012 | 86 | 2321 | 1203 |
| 2013 | 183 |  | 1218 |
| 2014 |  |  |  |
| 2015 |  |  |  |

Table 14.5.2: Number of recaptured cod in the period of 2003 to 2015 in different regions. Fjord (West) = NAFO divisions 1B-1F. Bank (West) = NAFO division 1D+1E. East Greenland = NAFO division 1F + ICES division 14.

| Recaptures |  |  |  |
| :--- | :---: | :---: | :---: |
|  | FJord (West) | Bank (West) | EAST Greenland |
| Fjord (West) | 438 | 14 | 2 |
| Bank (West) | 1 | 34 | 2 |
| East Greenland |  | 12 | 99 |
| Iceland | 3 | 21 | 125 |

### 14.13Figures



Figure. 14.1. Sampling location of spawning cod in Greenland and Iceland in the genetic project. The colours of the dots represent the blends of sample mean of the different spawning population: West offshore, Nuuk (inshore), East (Greenland and offshore Iceland) and Iceland inshore as signal intensities of green and red respectively. After Therkildsen et al. 2013.


Figure 14.2.1. Annual catch of cod in offshore West Greenland (NAFO subdivisions 1A-1E) used by the Working Group.


Figure 14.2.2.1: Annual distribution of total catches of Atlantic cod in West and East Greenland. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 14.2.2.1: Continued. Annual distribution of total catches of Atlantic cod in West and East Greenland. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 14.2.2.2: Distribution of Longline and Trawl catches of Atlantic cod in West and East Greenland 2015. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 14.2.3.1: Total length and age distributions of commercial cod catches in the West Greenland (NAFO 1A-1E) offshore fishery in 2015.


Figure 14.2.3.2: Catch-at-age in the West Greenland (NAFO 1A-1E) commercial fishery. Size of circles represents size of catch numbers.


Figure14.3.1.1. Greenland shrimp and fish survey 2008-2015. Abundance per $\mathrm{Km}^{2}$


Figure 14.3.1.2. Greenland shrimp and fish survey 2008-2015. Catch weight kg per $\mathrm{Km}^{2}$


Figure 14.3.1.3: Abundance index by age in NAFO 1A-1E combined. Size of circles represents index size of index.


Figure 14.3.1.4: West Greenland Shrimp and fish survey, 2015. Abundance index by length (cm) and area. Areas from north (top) to south (bottom) are: NAFO div. 1A; 1B+0A; 1C, 1D, 1E.


Figure 14.3.1.5. Abundance ( $\mathrm{no} / \mathbf{k m}^{2}$ ) pr. station of ages 1-10 in the years 2009-2015.


Figure 14.3.1.6: Total abundance indices by length in West Greenland shrimp and fish survey (NAFO 1A-1E).


Figure 14.3.1.7: Estimated SSB (tons) by NAFO subdivisions from the West Greenland Shrimp and Fish survey, 2015. Maturity taken from proportion mature by length as recorded on observer trips off East Greenland in 2007.


Figure 14.3.1.8: Estimated SSB (tons) by year from the West Greenland Shrimp and Fish survey (NAFO 1A-1E).


Figure 14.3.1.9: Composition of ogive state in females in survey in NAFO area 1D and 1E combined

Abundance distr. Atlantic cod, 2015


Figure 14.3.2.1 German survey, 2015. Abundance (num per km2) per haul.

Biomass distr. Atlantic cod, 2015


Figure 14.3.2.2 German survey, 2015. Biomass (kg per $\mathbf{k m} 2$ ) per haul.


Figure 14.3.2.3 German survey, Cod off Greenland. Abundance per age group and stratum in 2015. Strata 1-4 is West Greenland from north to south; strata 5-9 is East Greenland from south to north.


Figure 14.3.2.4 German survey, Cod off Greenland. Abundance indices for West Greenland (NAFO subdivisions 1C-1E).


Figure 14.3.2.5 German survey, Cod off Greenland. Biomass indices for West Greenland (NAFO subdivisions 1C-1E).

## 15 Cod in inshore waters of NAFO Subarea 1 (Greenland cod)

### 15.1 Stock definition

The cod found in Greenland is derived from four separate "stocks" that each is labelled by their spawning areas: I) offshore West Greenland waters; II) West Greenland fiords; III) offshore East Greenland and Icelandic waters and IV) inshore Icelandic waters (Therkildsen et al. 2013).

From 2012 the inshore component (West Greenland, NAFO Subarea 1) was assessed separately from all offshore components. From 2015 the offshore West Greenland (NAFO subdivisions 1A-E) and East Greenland (NAFO subdivision 1F and ICES Subarea 14) components was assessed separately. The Stock Annex provides more details on the stock identities including the references to the primary literature.

### 15.2 The fishery

Details on the historical development in the fisheries are provided in the stock annex.

### 15.2.1 The present fishery

The original TAC for the coastal fishery was set at 25,000 tons. In December 2,500 tons was added resulting in a total TAC of 27,500 t .
The coastal fishery took 25,272 tons +50 tons in Eastgreenland (Tasiilaq) in 2015, which is an increase of $38 \%$ compared to 2014 (table 15.2.1.1). The most important fishery is the poundnet fishery that takes place during summer followed by the fishery with jigs that takes place in autumn (table 15.2.1.2 and 15.2.1.3, figure 15.2.1.1). In 2014 and 2015, half of the total catch was taken by poundnets. This is a decrease compared to previous years where up to $3 / 4$ of the total catch where taken by poundnets (figure 15.2.1.1). Since 2012, jigs have become more dominant from $7 \%$ of the total catch in 2012 to 22 $\%$ of the total catch in 2015. Gillnets and longlines constitutes the rest of the total catch. The increase in the use of jigs are most likely caused by the increase in small dinghies that are participating in the fishery.

### 15.2.1.1 North Greenland (NAFO division 1A, subarea 1AX (Disco Bay))

The catches in North Greenland have gradually increased from 500 tons in 2012 to 4,100 tons in 2015 and catches now comprise $16 \%$ of the total catch (table 15.2.1.2). Never before in the time-series has the catches in this area been this high. The dominating gear are gillnets (Table 15.2.1.3) properly due to the fishing industry being concentrated on Greenland Halibut, therefore in this area, cod is mostly caught as bycatch in the gillnet settings for Greenland Halibut. Cod catches, especially in the southern part of Disco Bay, have however increased in recent years and directed cod fishery is presumed to be going on in this area (Figure 15.2.1.2).

The genetic study of the spawning cod in Disco Bay concluded that the cod are more similar to the offshore Westgreenlandic cod stock than the inshore cod stock (Therkildsen et al. 2013). The increasing numbers of cod in recent years and the facts that previous large year classes of Eastgrenlandic/Icelandic origin (1984 and 2003 YC) where not registered in Disco Bay (Storr-Paulsen et al. 2004 and Retzel \& Christensen 2016a) warrants for a more thorough investigation of the origin of the cod currently found in Disco Bay.

### 15.2.1.2 Midgreenland (NAFO divisions 1B, 1C and 1D)

Almost 21,000 tons cod were taken in this area in 2015 (table 15.2.1.1) corresponding to $83 \%$ of the total catches (table 15.2.1.2, figure 15.2.1.3). The last time catches in this area were this high was in 1989 and 1990 were the fishery fished on the 1984 YC . In all areas the dominating gear are poundnets followed by jigs which are especially used in area 1B and 1C (table 15.2.1.3, figure 15.2.1.3).

### 15.2.1.3 South Greenland (NAFO divisions 1E and 1F)

The catches in South Greenland have the last couple of years gradually declined to 230 tons in 2015 corresponding to $1 \%$ of the total inshore catch (table 15.2.1.1 and 15.2.1.2, figure 15.2.1.3). Never before in the time-series has a pattern with very little catches in South Greenland simultaneously with high catches in the other areas been observed. The last time there were catches over 25,000 tons inshore was in 1989 and 1990 where a large 1984 YC supported the fishery. This YC was of offshore East Greenland/Icelandic origin and distributed offshore and inshore along the coastline and the inshore fishery in South Greenland was high as the fish migrated inshore (Storr-Paulsen et al. 2004). The inshore cod stock is believed to be distributed from Midgreenland and northwards as there are no significant spawning taking place in South Greenland (Retzel \& Hedeholm 2012), and the fishery here is therefore very much depending on offshore fish migrating inshore. Currently, there is a 2010 YC of average size distributed inshore and offshore in Midgreenland, while a 2009 YC of considerable size is distributed offshore in South Greenland (Retzel 2016a, Retzel \& Christensen 2016a). The offshore fish of these year classes are believed to be of East Greenland origin. For some reason the 2009 YC do not migrate inshore in South Greenland.

### 15.2.1.4East Greenland (ICES Subdivision 14b)

A very small amount ( 50 tons, table 15.2.1.1) of the inshore quota is fished in East Greenland with jigs in the Tasiilaq area (Figure 15.2.1.2 and 15.2.1.3). There are no length measurements from this fishery and the fish here are not presumed to belong to the inshore West Greenland cod stock. These fish are therefore not included the overall calculations of Catch and Weight at age.

### 15.2.2 Length, weight and age distributions

In 2015 the Greenland inshore length frequencies were measured from 116 inshore samples (21 854 cod measured).
Several YC were caught in the inshore fishery in 2015, and ages 4-7 (YC 2008-2011) comprised the catches in 2015, with the 2010 YC dominating the catches followed by the 2009 YC (table 15.2.2.1, figure 15.2.2.1 and 15.2.2.2). Mean length in catches have increased from 53 cm in 2010-2013 to 58 in 2014 and 57 cm in 2015. In 2013 and 2014, the 2009 YC was the dominating YC in the catches (table 15.2.2.1). However, from the offshore Shrimp and Fish survey (Retzel \& Christensen 2016a, Retzel 2016a) it is documented that the 2009 YC is migrating southwards along the coast, and the 2010 YC is the dominating YC in the offshore area of NAFO 1C and 1D. Furthermore, as the 2009 YC increases their length with the growing age they do no longer get caught in the fishing gear used in the inshore fishery.

### 15.2.3 Catch Curve Analysis

A Catch Curve Analysis (CAA) was performed on the catch-at-age data for each YC from 1973-2007 and Z was calculated for the ages 4-8 (table 15.2.2.3). For the YC 1990-

1997 point/years were missing in the data as there was no sampling of the fishery in 1998 and 2001. In general, the CAA performed well with high $R^{2}$. Overall $Z$ was high in most years. The relatively strong YC that produced catches above 10,000 tons (1984, 1985 and 2003) had $Z$ values around 1.5. It is however, not easy to disentangle the effect of fishery and natural mortality, as the latter is subject to migration and a selective fishery with unknown age specific availability. Some input from the coastal and offshore region in the younger YC (ages 4-6) must be expected, but these fish tend to undertake spawning migrations when they reach maturity (ICES 2012). It is unknown in what quantity this contributes to the fishery. The availability changes with age due to the nature of the fishery, which is mainly conducted with poundnets in shallow water (table 15.2.1.3). Older fish (>8) tend to migrate away from the shallow water, and become unavailable to the fishery. This inflates the natural mortality but it has not been quantified. Hence, the combined effect of migration and age availability in the poundnet fishery causes the natural mortality to increase, but it is unknown to what extent. Nevertheless, Z remains high, but is likely an overestimate.

### 15.3 Survey

### 15.3.1 Results of the West Greenland gillnet survey

The numbers of valid net settings in 2015 was 44 in NAFO 1B and 59 in NAFO 1D (Table 15.3.1.1). Area and site-specific catch rates can be seen in figure. 15.3.1.1.

In 1B age 2 and 3 fish ( 2012 and $2013 \mathrm{YC}^{\prime}$ s), which the survey mainly targets, appear to be small cohorts, and are smaller than the time-series mean (Table 15.3.1.2, figure. 15.3.1.2). The same pattern was seen in 2014 with the YC 2011 (as 3 years old) and the 2012 YC (as 2 years old). Since the 2009 YC, no strong YC has been documented in 1B and recruitment is low compared with earlier years (Table 15.3.1.2, Figure. 15.3.1.3). Overall, the NAFO 1B index (including all age groups) increased with 62 \% from 2014 to 2015, caused by increased numbers of age 4,5 and 6 year olds (YC 2009-2011).

The 2015 catches in NAFO 1D were dominated by 3 and 4 years old cod (2011 and 2012 YC's, Table 15.3.1.2). Catch rates of 2012 YC in 2015 was the second highest in the timeseries. The 2009 YC was not an outstanding cohort at age 2 and 3 in 2011 and 2012 in 1D, but the index increased for age 4 in 2013 and was one of the highest recorded indices in 2013, and at age 5 and 6 in 2014 and 2015, respectively. The overall index for NAFO 1D (including all age groups) is the second highest in the time-series, and increased by $17 \%$ in 2015 compared to 2014 (Table 15.3.1.2).

Combining the two NAFO (1B and 1D) divisions in a joint index shows an overall decline of 49 \% in total index for all ages from 2013-2014, which was the lowest index since 2006. Although the index increased by 36 \% between 2014 and 2015 the index is still low compared to 2010-2013 (Figure. 15.3.1.4). The overall trend for the divisions is driven by the development in NAFO 1B, where the catch rates and index values are normally higher than 1D (Table 15.3.1.2, Figure. 15.2.1.3). However, in 2014 the total index was higher in 1D, caused by the index being higher for especially ages 3 and 4 ( 2011 and 2010 YC) in 1D. In 2015, age 3 and 4 (2011 and 2012 YC) dominated the index.

The combined index for 1B and 1D for age 2 and 3 jointly in 2014 and 2015 has decreased by $61 \%$ compared to the average of the preceding four years (2010-2013).

### 15.3.2 Surveys in North Greenland (Disco Bay)

Currently two surveys are conducted in Disco Bay 1) a trawl survey targeting shrimp (part of the Greenland shrimp and fish survey covering all offshore waters in Greenland) and 2) Gillnet survey targeting Greenland Halibut. Since 2011 increasing amount of cod have been caught in these surveys, especially the 2009 YC. The results are not used in the assessment and further details and results of these surveys can be found in working document nr 18 (Retzel \& Christensen 2016b).

### 15.4 Information on spawning

In 2011 a survey was conducted in spring in order to investigate the extent of spawning in fjords not traditionally surveyed. The results show that spawning occurs in most fjords and is especially pronounced between Sisimiut (NAFO 1B) and Paamiut (NAFO 1E). Further information is provided in the stock annex.

### 15.5 Tagging experiments

A total of 17304 cod have been tagged in different regions of Greenland in the period of 2003-2015 (table 15.5.1). 4282 cod have been tagged in the inshore area in West Greenland primarily in NAFO 1B, 1D and 1F. Largest numbers of tagged fish occurred 2003-2009. Since 2009 limited amount of cod have been tagged inshore.

Innshore recaptures are found almost exclusively in the same ford as tagged (table 15.5.2). Only one fish have been recaptured on the bank in West Greenland and 3 have been recaptured in Iceland. All three recaptures from Iceland were tagged in South Greenland (NAFO 1F).

### 15.6 State of the stock

There have been several years of steady and relatively high recruitment for ten years. The past two years however recruitment has been low and is below average for the whole time-series. Index of older ages (4+) are above average due to the high recruitment but is decreasing. Catches have risen since 2000 and are in 2015 at their highest level in 25 years. Combining these trends suggests a recent increase in exploitation rate.
Several year classes are in the catches, and the large 2009 YC is now the second largest YC in the fishery at age 6 . The 2010 YC is dominating the fishery at age 5 and this YC is considered smaller than the 2009 YC . The fishery in concentrated on $4-6 \mathrm{yr}$ old and after the 2009 YC has gone through the fishery no new incoming yearclasses of the same size has been observed, and recruitment of the 2012 and 2013 YC are considered very low. Spawning has been documented in most fjords on the west coast, with key areas in NAFO 1B and 1D. Hence the overall state of the stock is considered as stable, but the lack of incoming large yearclasses is cause for concern.

### 15.7 Implemented management measures for 2015

Until 2009 the inshore fishery was unregulated by a TAC. The TAC in 2009-2014 can be seen in figure 15.1.1.2. The TAC for 2015 is set at 26000 t . No other management measures have been taken.

### 15.8 Management plan

No management plan currently exists for the inshore cod stock.

### 15.9 Management considerations

When managing this stock, it should be taken into consideration that the inshore cod tend to form very dense spawning aggregations in limited areas. It could be considered to limit the fishery in certain areas or certain periods, especially if the stock shows a declining trend. These areas include specifically certain areas in the Nuuk and Sisimiut fjord systems.

Genetic and tagging results indicate limited migration between fjords and management should therefore ensure that not all catches are taken in a limited area. This is especially important in areas that are considered to have maintained the stocks in periods of overall stock decline in Greenland (i.e. Nuuk and Sisimiut fjords).

The fishery in this region is a mixed-stock fishery including other Greenland cod stocks (south and east Greenland cod, as well as offshore west Greenland cod) and Iceland cod. No operational procedure exists to evaluate the proportional contribution in the catches.

### 15.10 Basis for advice

The survey index in a given year was related to the catch in the next year (Figure D.1.1). The advice is then based on the survey index multiplied by a factor. The validity of this approach rests on a number of assumptions. Among others, the fishery has been at a stable sustainable level (ideally the same across years). Based on model outputs and catch curves (Hedeholm and Post, 2015) this seems to be a reasonable assumption, at least during the last 15 years. Some years in the 1980s did not follow the overall trend, and were most likely subjected to a very high fishing intensity and a very high offshore input to the fishery, and these years are therefore excluded from the regression analyses. The fish enter the fishery at age 4 . Accordingly the survey index of ages $3-8$ was used to generate advice.


Figure D.1.1. Survey index of 3-8 year olds vs. the catch the following year. $r^{2}=0.76$. Based on data from 1991-2014. Points are labelled by survey year.

Given that this approach is based on variable data a precautionary approach should be taken. So rather than having the regression pass through the origin, the intercept with the $x$-axis is set at a survey index value of 50 and slope is 37.9 . The survey tends to vary considerably between years, and to avoid having the advice fluctuate accordingly the average of the last two years survey index values were used when calculating the catch advice. Consequently, the advice is generated as follows:
$\mathrm{C}_{\mathrm{y}+1}=37.9$ * ( $\mathrm{U}_{3-8 y}-\mathrm{U}_{\text {trigger }}$ ) (1)
where $\mathrm{U}_{3-8 \mathrm{y}}$ is the combined survey value for ages 3-8 and $U_{\text {trigger }}$ is 50 .
The above described procedure agreed upon by the benchmark was considered less than optimal by ADG due to the possibility of fluctuating advice. The ADG concluded that there was no trend in the adult survey index (age 3-8) and therefore the same advice as last year was given. The same advice has now been given for three years and the advice is based on a $20 \%$ increase of average catch in 2010, 2011 and 2012.

The NWWG concluded that as the survey index of the adult part of the stock is still fluctuating without trend same advice as last year is still valid. However using the DLS 3.2 approach could also be implemented with the survey index for the $2+3$ year as a stock indicator and basis for the advice. Advice is given 2 yrs later (2017) than the survey results (2015), and given a high degree of internal consistency in the survey between the $2+3$ year old compared with the $4+5$ year old two years later (figure 15.10.1 and 15.10.2). As the cod enter the fishery at age 4 the survey index for the $2+3 \mathrm{yr}$ old is considered a good candidate for the stock indicator.

### 15.11References

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### 15.12Tables

Table 15.2.1.1. Cod catches (t) divided into NAFO divisions, caught in the inshore fishery (19111993: Horsted 2000, 1994-2006: ICES 2007, Statistic Greenland, 2007-present: Greenland Fisheries License Control). ICES 14.b=inshore East Greenland.

| NAFO DIVISIONS |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 A | 1 B | 1 C | 1D | 1E | 1F | Unknown NAFO DIV | Total <br> WestGreentand | $\begin{aligned} & \text { ICES } \\ & \text { XIVB } \end{aligned}$ |
| 1911 |  |  |  | 19 |  |  |  | 19 |  |
| 1912 |  |  |  | 5 |  |  |  | 5 |  |
| 1913 |  |  |  | 66 |  |  |  | 66 |  |
| 1914 |  |  |  | 60 |  |  |  | 60 |  |
| 1915 |  | 47 | 6 | 45 |  |  |  | 98 |  |
| 1916 |  | 66 | 24 | 103 |  |  |  | 193 |  |
| 1917 |  | 67 | 28 | 59 |  |  |  | 154 |  |
| 1918 |  | 106 | 26 | 140 |  | 169 |  | 441 |  |
| 1919 |  | 39 | 37 | 140 | 148 | 137 |  | 501 |  |
| 1920 |  | 117 | 32 | 187 | 23 | 95 |  | 454 |  |
| 1921 |  | 116 | 92 | 97 | 7 | 196 |  | 508 |  |
| 1922 |  | 82 | 178 | 144 | 40 | 158 |  | 602 |  |
| 1923 |  | 120 | 116 | 147 | 0 | 307 |  | 690 |  |
| 1924 |  | 131 | 223 | 221 | 1 | 267 |  | 843 |  |
| 1925 |  | 122 | 371 | 318 | 45 | 168 |  | 1024 |  |
| 1926 |  | 97 | 785 | 673 | 170 | 499 |  | 2224 |  |
| 1927 |  | 282 | 974 | 982 | 305 | 1027 |  | 3570 |  |
| 1928 |  | 426 | 888 | 1153 | 497 | 1199 |  | 4163 |  |
| 1929 |  | 1479 | 1572 | 1335 | 642 | 2052 |  | 7080 |  |
| 1930 | 137 | 2208 | 2326 | 1681 | 994 | 2312 |  | 9658 |  |
| 1931 | 315 | 1905 | 2026 | 1520 | 835 | 2453 |  | 9054 |  |
| 1932 | 358 | 1713 | 2130 | 1042 | 731 | 3258 |  | 9232 |  |
| 1933 | 304 | 1799 | 1743 | 1148 | 948 | 2296 |  | 8238 |  |
| 1934 | 451 | 2080 | 1473 | 652 | 921 | 3591 |  | 9168 |  |
| 1935 | 524 | 1870 | 1277 | 769 | 670 | 2466 |  | 7576 |  |
| 1936 | 329 | 2039 | 1199 | 705 | 717 | 2185 |  | 7174 |  |
| 1937 | 135 | 1982 | 1433 | 854 | 496 | 2061 |  | 6961 |  |
| 1938 | 258 | 1743 | 1406 | 703 | 347 | 1035 |  | 5492 |  |
| 1939 | 416 | 2256 | 1732 | 896 | 431 | 1430 |  | 7161 |  |
| 1940 | 482 | 2478 | 1600 | 1061 | 646 | 1759 |  | 8026 |  |
| 1941 | 636 | 3229 | 1473 | 823 | 593 | 1868 |  | 8622 |  |
| 1942 | 879 | 3831 | 2249 | 1332 | 1003 | 2733 |  | 12027 |  |
| 1943 | 1507 | 5056 | 2016 | 1240 | 1134 | 2073 |  | 13026 |  |
| 1944 | 1795 | 4322 | 2355 | 1547 | 1198 | 2168 |  | 13385 |  |
| 1945 | 1585 | 4987 | 2844 | 1207 | 1474 | 2192 |  | 14289 |  |
| 1946 | 1889 | 5210 | 2871 | 1438 | 1139 | 2715 |  | 15262 |  |
| 1947 | 1573 | 5261 | 3323 | 2096 | 1658 | 4118 |  | 18029 |  |
| 1948 | 1130 | 5660 | 3756 | 1657 | 1652 | 4820 |  | 18675 |  |
| 1949 | 1403 | 4580 | 3666 | 2110 | 2151 | 3140 |  | 17050 |  |
| 1950 | 1657 | 6358 | 4140 | 2357 | 2278 | 4383 |  | 21173 |  |
| 1951 | 1277 | 5322 | 3324 | 2571 | 2101 | 3605 |  | 18200 |  |
| 1952 | 646 | 4443 | 2906 | 2437 | 2216 | 4078 |  | 16726 |  |
| 1953 | 1092 | 5030 | 3662 | 5513 | 3093 | 4261 |  | 22651 |  |


| NAFO DIVISIONS |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 A | 1 B | 1 C | 1 D | 1 E | 1 F | Unknown | Total | ICES |
|  |  |  |  |  |  |  | NAFO DIV | WestGreenland | XIVв |
| 1954 | 950 | 6164 | 3118 | 3275 | 1773 | 3418 |  | 18698 |  |
| 1955 | 591 | 5523 | 3225 | 4061 | 2773 | 3614 |  | 19787 |  |
| 1956 | 475 | 5373 | 3175 | 5127 | 3292 | 3586 |  | 21028 |  |
| 1957 | 277 | 6146 | 3282 | 5257 | 4380 | 5251 |  | 24593 |  |
| 1958 | 19 | 6178 | 3724 | 5456 | 3975 | 6450 |  | 25802 |  |
| $1959$ | 237 | 6404 | 5590 | 5009 | 3767 | 6570 |  | 27577 |  |
| 1960 | 188 | 6741 | 6230 | 3614 | 3626 | 6610 |  | 27009 |  |
| 1961 | 601 | 6569 | 6726 | 4178 | 6182 | 9709 |  | 33965 |  |
| 1962 | 315 | 7809 | 6269 | 3824 | 5638 | 11525 |  | 35380 |  |
| 1963 | 295 | 4877 | 3178 | 2804 | 3078 | 9037 |  | 23269 |  |
| 1964 | 275 | 3311 | 2447 | 8766 | 2206 | 4981 |  | 21986 |  |
| 1965 | 325 | 5209 | 4818 | 6046 | 2477 | 5447 |  | 24322 |  |
| 1966 | 483 | 8738 | 5669 | 7022 | 2335 | 4799 |  | 29046 |  |
| 1967 | 310 | 5658 | 6248 | 6747 | 2429 | 6132 |  | 27524 |  |
| 1968 | 142 | 1669 | 2738 | 6123 | 2837 | 7207 |  | 20716 |  |
| 1969 | 57 | 1767 | 4287 | 7540 | 2017 | 5568 |  | 21236 |  |
| 1970 | 136 | 1469 | 2219 | 3661 | 2424 | 5654 |  | 15563 |  |
| 1971 | 255 | 1807 | 2011 | 3802 | 1698 | 3933 |  | 13506 |  |
| 1972 | 263 | 1855 | 3328 | 3973 | 1533 | 3696 |  | 14648 |  |
| 1973 | 158 | 1362 | 1225 | 3682 | 1614 | 1581 |  | 9622 |  |
| $1974$ | 454 | 926 | 1449 | 2588 | 1628 | 1593 |  | 8638 |  |
| 1975 | 216 | 1038 | 1930 | 1269 | 964 | 1140 |  | 6557 |  |
| 1976 | 204 | 644 | 1224 | 904 | 1367 | 831 |  | 5174 |  |
| 1977 | 216 | 580 | 2505 | 2946 | 3521 | 4231 |  | 13999 |  |
| 1978 | 348 | 1587 | 3244 | 2614 | 4642 | 7244 |  | 19679 |  |
| 1979 | 433 | 1768 | 2201 | 6378 | 9609 | 15201 |  | 35590 |  |
| 1980 | 719 | 2303 | 2269 | 7781 | 10647 | 14852 |  | 38571 |  |
| 1981 | 281 | 2810 | 3599 | 6119 | 7711 | 11505 | 7678 | 39703 |  |
| 1982 | 206 | 2448 | 3176 | 7186 | 4536 | 3621 | 5491 | 26664 |  |
| 1983 | 148 | 2803 | 3640 | 7430 | 5016 | 2500 | 7205 | 28742 |  |
| 1984 | 175 | 3908 | 1889 | 5414 | 1149 | 1333 | 6090 | 19958 |  |
| $1985$ | 149 | 2936 | 957 | 1976 | 1178 | 1245 |  | 8441 |  |
| $1986$ | 76 | $1038$ | 255 | $1209$ | $1456$ | 1268 |  | 5302 |  |
| 1987 | 77 | 2366 | 423 | 6407 | 3602 | 1326 | 403 | 14604 |  |
| 1988 | 333 | 6294 | 1342 | 2992 | 3346 | 4484 |  | 18791 |  |
| 1989 | 634 | 8491 | 5671 | 8212 | 10845 | 4676 |  | 38529 |  |
| 1990 | 476 | 9857 | 1482 | 9826 | 1917 | 5241 |  | 28799 |  |
| 1991 | 876 | 8641 | 917 | 2782 | 1089 | 4007 |  | 18312 |  |
| 1992 | 695 | 2710 | 563 | 1070 | 239 | 450 |  | 5727 |  |
| 1993 | 333 | 327 | 168 | 970 | 19 | 109 |  | 1926 |  |
| 1994 | 209 | 332 | 589 | 914 | 11 | 62 |  | 2117 |  |
| 1995 | 53 | 521 | 710 | 332 | 4 | 81 |  | 1701 |  |
| 1996 | 41 | 211 | 471 | 164 | 11 | 46 |  | 944 |  |
| 1997 | 18 | 446 | 198 | 99 | 13 | 130 | 282 | 1186 |  |
| 1998 | 9 | 118 | 79 | 78 | 0 | 38 |  | 322 |  |


| NAFO DIVISIONS |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 A | 1 B | 1 C | 1 D | 1 E | 1 F | Unknown | Total | ICES |
|  |  |  |  |  |  |  | NAFO DIV | WestGreenland | XIVB |
| 1999 | 68 | 142 | 55 | 336 | 8 | 4 |  | 613 |  |
| 2000 | 154 | 266 | 0 | 332 | 0 | 12 |  | 764 |  |
| 2001 | 117 | 1183 | 245 | 54 | 0 | 81 |  | 1680 |  |
| 2002 | 263 | 1803 | 505 | 214 | 24 | 813 |  | 3622 |  |
| 2003 | 1109 | 1522 | 334 | 274 | 3 | 479 | 1494 | 5215 |  |
| 2004 | 535 | 1316 | 242 | 116 | 47 | 84 | 2608 | 4948 |  |
| 2005 | 650 | 2351 | 1137 | 1162 | 278 | 382 | 83 | 6043 |  |
| 2006 | 922 | 1682 | 577 | 943 | 630 | 1461 | 1173 | 7388 |  |
| 2007 | 416 | 2547 | 1195 | 1842 | 659 | 4391 |  | 11050 | 42 |
| 2008 | 870 | 3066 | 1539 | 3172 | 225 | 1133 |  | 10005 | 6 |
| 2009 | 325 | 1288 | 1189 | 2009 | 1142 | 1581 |  | 7534 | 2 |
| 2010 | 559 | 2990 | 1607 | 1795 | 1458 | 859 |  | 9268 | 2 |
| 2011 | 567 | 2364 | 2850 | 2905 | 1274 | 1047 |  | 11007 | 0 |
| 2012 | 546 | 1376 | 2061 | 4375 | 1989 | 325 |  | 10672 | 0.02 |
| 2013 | 1506 | 2552 | 2784 | 4711 | 1450 | 198 |  | 13202 | 35 |
| 2014 | 3084 | 6142 | 3710 | 4629 | 684 | 82 |  | 18331 | 38 |
| 2015 | 4088 | 7912 | 6426 | 6613 | 117 | 115 |  | 25272 | 50 |

Table 15.2.1.2: Catches ( $t$ ) divided into month and NAFO Divisions, caught by the coastal fisheries.

| NAFO | JAN | FEB | MAR | APR | MAY | JUN | JUL | AUG | SEP | Oct | NOV | DEC | TOTAL | \% |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1AUM | 3 | 9 | 13 | 2 |  | 1 | 0.4 | 0.2 | 37 | 29 | 29 | 1 | 123 | $0.5 \%$ |
| 1AUP |  |  |  |  |  |  | 0.01 | 0.01 | 0.1 | 0.003 |  |  | 0.1 | $0.001 \%$ |
| 1AX | 173 | 183 | 103 | 137 | 63 | 124 | 413 | 565 | 507 | 873 | 671 | 155 | 3965 | $16 \%$ |
| 1B | 86 | 86 | 317 | 282 | 706 | 873 | 1535 | 1005 | 894 | 996 | 980 | 153 | 7912 | $31 \%$ |
| 1C | 197 | 78 | 56 | 82 | 199 | 1159 | 1182 | 721 | 834 | 824 | 599 | 495 | 6426 | $25 \%$ |
| 1D | 240 | 107 | 180 | 314 | 543 | 1299 | 1516 | 644 | 770 | 533 | 274 | 193 | 6613 | $26 \%$ |
| 1E | 0.4 | 1 | 0.1 | 2 | 1 | 11 | 56 | 39 | 4 | 3 | 0.4 | 0.02 | 117 | $0.5 \%$ |
| 1F | 0.4 | 1 |  | 0.4 | 0.1 | 18 | 1 | 2 | 7 | 21 | 59 | 4 | 115 | $0.5 \%$ |
| Total | 698 | 463 | 670 | 819 | 1511 | 3486 | 4703 | 2977 | 3052 | 3279 | 2612 | 1002 | 25272 |  |
| \% | $3 \%$ | $2 \%$ | $3 \%$ | $3 \%$ | $6 \%$ | $14 \%$ | $19 \%$ | $12 \%$ | $12 \%$ | $13 \%$ | $10 \%$ | $4 \%$ |  |  |
| ICES |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| XIVb |  |  |  |  |  |  |  |  | 24 | 25 |  |  | 50 |  |

Table 15.2.1.3: Landings (\%) divided into month and gear and NAFO Divisions and gear.

| GEAR/MONTH | JAN | FEB | MAR | APR | MAY | JUN | JUL | AUG | SEP | OCt | NOV | DEC | TOTAL |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Poundnet | $0.003 \%$ |  |  | $1 \%$ | $5 \%$ | $13 \%$ | $14 \%$ | $5 \%$ | $4 \%$ | $5 \%$ | $4 \%$ | $0.2 \%$ | $50 \%$ |
| Gillnet | $1 \%$ | $1 \%$ | $2 \%$ | $1 \%$ | $0.5 \%$ | $0.1 \%$ | $0.1 \%$ | $0.2 \%$ | $2 \%$ | $3 \%$ | $3 \%$ | $1 \%$ | $15 \%$ |
| Jig | $0.03 \%$ | $0.1 \%$ | $0.3 \%$ | $1 \%$ | $0.2 \%$ | $1 \%$ | $4 \%$ | $6 \%$ | $6 \%$ | $3 \%$ | $1 \%$ | $0.3 \%$ | $22 \%$ |
| Longline | $1 \%$ | $1 \%$ | $1 \%$ | $1 \%$ | $0.2 \%$ | $1 \%$ | $0 \%$ | $0.5 \%$ | $1 \%$ | $2 \%$ | $2 \%$ | $2 \%$ | $13 \%$ |
| Total | $3 \%$ | $2 \%$ | $3 \%$ | $3 \%$ | $6 \%$ | $14 \%$ | $19 \%$ | $12 \%$ | $12 \%$ | $13 \%$ | $10 \%$ | $4 \%$ | $100 \%$ |


| GEAR/NAFO | IAUM | IAUP | IAX | 1B | IC | 1D | IE | IF | ICES <br> XIVB | TOTAL |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Poundnet |  |  | $4 \%$ | $16 \%$ | $12 \%$ | $18 \%$ | $0.4 \%$ | $0.1 \%$ |  |  |
| Gillnet | $0.1 \%$ |  | $6 \%$ | $6 \%$ | $1 \%$ | $2 \%$ | $0.004 \%$ | $0.1 \%$ | $15 \%$ |  |
| Jig | $0.05 \%$ |  | $4 \%$ | $8 \%$ | $7 \%$ | $3 \%$ | $0.1 \%$ | $0.1 \%$ | $0.2 \%$ | $22 \%$ |
| Longline | $0.3 \%$ | $0.001 \%$ | $2 \%$ | $1 \%$ | $6 \%$ | $3 \%$ | $0.03 \%$ | $0.2 \%$ | $13 \%$ |  |
| Total | $0.5 \%$ | $0.001 \%$ | $16 \%$ | $31 \%$ | $25 \%$ | $26 \%$ | $0.5 \%$ | $0.5 \%$ | $0.2 \%$ | $100 \%$ |

Table 15.2.2.1. Estimated catches in numbers ('000) at age, and total catch by year (t).

|  | AGE |  |  |  |  |  |  |  | Tonnes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | LANDED |
| 1976 | 2508 | 924 | 556 | 287 | 38 | 31 | 11 | 7 | 5174 |
| 1977 | 467 | 5437 | 1100 | 883 | 179 | 7 | 142 | 46 | 13999 |
| 1978 | 97 | 1262 | 9904 | 132 | 68 | 7 | 3 |  | 19679 |
| 1979 | 323 | 2297 | 2380 | 8281 | 170 | 96 | 4 | 14 | 35590 |
| 1980 | 4343 | 4334 | 1646 | 806 | 6492 | 106 | 29 | 37 | 38571 |
| 1981 | 87 | 15793 | 5225 | 725 | 499 | 2906 | 61 | 17 | 39703 |
| 1982 | 3013 | 1587 | 6309 | 1545 | 798 | 152 | 610 | 154 | 26664 |
| 1983 | 229 | 16877 | 1381 | 4352 | 368 | 139 | 65 | 75 | 28742 |
| 1984 | 520 | 4451 | 9269 | 346 | 634 | 18 | 42 | 12 | 19958 |
| 1985 | 5 | 2400 | 1028 | 2229 | 196 | 363 | 14 | 78 | 8441 |
| 1986 | 286 | 178 | 896 | 460 | 721 | 16 | 102 | 38 | 5302 |
| 1987 | 5503 | 1334 | 228 | 710 | 340 | 1084 | 46 | 265 | 14604 |
| 1988 | 419 | 15588 | 150 | 51 | 39 | 90 | 161 | 12 | 18791 |
| 1989 | 15 | 5962 | 23956 | 271 | 46 | 2 | 93 | 176 | 38529 |
| 1990 | 212 | 2997 | 15403 | 6732 | 33 | 11 | 7 | 16 | 28799 |
| 1991 | 124 | 6022 | 4910 | 5695 | 330 | 0 |  |  | 18312 |
| 1992 | 8 | 2408 | 2344 | 452 | 139 | 46 | 13 | 5 | 5727 |
| 1993 | 28 | 661 | 575 | 206 | 34 | 41 | 10 | 7 | 1926 |
| 1994 | 22 | 1468 | 342 | 62 | 45 | 8 | 11 | 1 | 2117 |
| 1995 | 1 | 834 | 773 | 37 | 5 | 0 | 0 |  | 1701 |
| 1996 | 2 | 165 | 362 | 130 | 25 | 3 | 1 | 0 | 944 |
| 1997 | 1 | 397 | 311 | 179 | 31 | 0 |  |  | 1186 |
| 1998* |  |  |  |  |  |  |  |  | 322 |
| 1999 | 87 | 465 | 105 | 1 | 0 | 0 |  |  | 613 |
| 2000 | 4 | 228 | 336 | 7 | 0 | 0 |  |  | 764 |
| 2001* |  |  |  |  |  |  |  |  | 1680 |
| 2002 | 532 | 2243 | 657 | 29 | 9 | 1 | 0 | 0 | 3622 |
| 2003 | 152 | 581 | 1547 | 258 | 51 | 16 | 15 | 11 | 5215 |
| 2004 | 530 | 1669 | 1095 | 228 | 37 | 3 |  |  | 4948 |
| 2005 | 1392 | 2408 | 944 | 186 | 36 | 10 | 4 | 0 | 6043 |
| 2006 | 4256 | 3363 | 680 | 22 | 0 | 0 | 0 |  | 7388 |
| 2007 | 1944 | 7910 | 1010 | 116 | 38 | 13 | 8 | 4 | 11050 |
| 2008 | 1176 | 5012 | 2793 | 319 | 36 | 6 | 2 |  | 10005 |
| 2009 | 487 | 3540 | 2372 | 194 | 13 | 3 | 0 | 4 | 7534 |
| 2010 | 301 | 1091 | 2475 | 1524 | 141 | 32 | 21 | 27 | 9268 |
| 2011 | 129 | 2929 | 2567 | 1480 | 255 | 90 | 12 | 7 | 11007 |
| 2012 | 735 | 1725 | 2681 | 850 | 182 | 21 | 13 | 13 | 10672 |
| 2013 | 143 | 3806 | 2477 | 1083 | 361 | 115 | 67 | 9 | 13202 |
| 2014 | 40 | 1389 | 4024 | 2292 | 328 | 168 | 103 | 52 | 18331 |
| 2015 | 20 | 2007 | 5683 | 3010 | 1338 | 133 | 9 | 8 | 25272 |

Table 15.2.2.2. West Greenland inshore cod. Estimated weight at age (kg).


Table 15.2.2.3. West Greenland inshore cod. Catch curve analysis. YearClass mortalities at ages 4-8 estimated from commercial catch-at-age data. * few data due to years (1998 and 2001) with no sampling. Yellow highlights strong YearClasses.

| YearClass | Z (4-8) | R2 |
| :---: | :---: | :---: |
| 1973 | 0.17 | 0.31 |
| 1974 | 0.58 | 0.78 |
| 1975 | 0.63 | 0.85 |
| 1976 | 1.36 | 0.85 |
| 1977 | 0.98 | 0.95 |
| 1978 | 1.11 | 0.89 |
| 1979 | 0.8 | 0.87 |
| 1980 | 0.89 | 0.96 |
| 1981 | 1.73 | 0.89 |
| 1982 | 0.72 | 0.87 |
| 1983 | 1.59 | 0.85 |
| 1984 | 1.59 | 0.87 |
| 1985 | 1.47 | 0.78 |
| 1986 | 1.68 | 0.91 |
| 1987 | 2.14 | 0.95 |
| 1988 | 1.81 | 0.97 |
| 1989 | 2.22 | 0.82 |
| 1990* | 1.34 | 0.97 |
| 1991* |  |  |
| 1992* |  |  |
| 1993* |  |  |
| 1994* |  |  |
| 1995* | 1.03 | 0.87 |
| 1996* | 0.94 | 0.79 |
| 1997* | 1.45 | 0.98 |
| 1998 | 1.43 | 0.94 |
| 1999 | 1.06 | 0.86 |
| 2000 | 1.46 | 0.87 |
| 2001 | 1.63 | 0.98 |
| 2002 | 1.37 | 0.86 |
| 2003 | 1.19 | 0.89 |
| 2004 | 1.32 | 0.91 |
| 2005 | 0.95 | 0.91 |
| 2006 | 0.57 | 0.74 |
| 2007 | 0.83 | 0.95 |

Table 15.3.1.1: Survey effort in the Greenland Inshore Gillnet survey (nos. of valid net settings).

| Division | 1 B | 1D | 1F | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1985 | 3 | 38 | 27 | 68 |
| 1986 | 26 | 22 | 23 | 71 |
| 1987 | 24 | 27 | 26 | 77 |
| 1988 | 21 | 24 | 24 | 69 |
| 1989 | 28 | 19 | 32 | 79 |
| 1990 | 18 | 21 | 18 | 57 |
| 1991 | 23 | 24 | 20 | 67 |
| 1992 | 27 | 29 | 23 | 79 |
| 1993 | 23 | 25 | 19 | 67 |
| 1994 | 20 | 29 | 17 | 66 |
| $1995$ | 24 | 21 | 20 | 65 |
| 1996 | 26 | 25 | - | 51 |
| 1997 | 20 | 23 | - | 43 |
| 1998 | 24 | 26 | 22 | 72 |
| 1999 | - | 24 | - | 24 |
| 2000 | - | 27 | 20 | 47 |
| 2001 | - | - | - | - |
| 2002 | 21 | 20 | - | 41 |
| 2003 | 33 | 27 | - | 60 |
| 2004 | 27 | 31 | - | 58 |
| 2005 | 25 | 28 | - | 53 |
| 2006 | 45 | 51 | - | 96 |
| 2007 | 52 | - | 39 | 91 |
| 2008 | - | 58 | 60 | 118 |
| 2009 | - | 58 | 18 | 76 |
| 2010 | 66 | 52 | - | 118 |
| 2011 | 57 | 44 | - | 101 |
| 2012 | 54 | 52 | - | 106 |
| 2013 | 58 | 52 | - | 110 |
| 2014 | 60 | 41 | - | 101 |
| 2015 | 59 | 44 | - | 103 |

Table 15.3.1.2: NAFO Div. 1B. Cod abundance indices (numbers of cod caught per 100 hours net settings) by age in the West Greenland inshore gillnet survey. na= data not available.

| Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | All |
| 1985 | 26 | 23 | 0 | 6 | 0 | 0 | 0 | 0 | 54 |
| 1986 | 4 | 245 | 16 | 8 | 2 | 2 | 0 | 0 | 278 |
| 1987 | 0 | 122 | 233 | 25 | 1 | 0 | 0 | 0 | 381 |
| 1988 | 0 | 33 | 130 | 111 | 2 | 0 | 0 | 0 | 276 |
| 1989 | 1 | 110 | 83 | 57 | 32 | 1 | 0 | 0 | 283 |
| 1990 | 0 | 109 | 108 | 62 | 53 | 12 | 0 | 0 | 344 |
| 1991 | 0 | 3 | 131 | 53 | 11 | 3 | 0 | 0 | 202 |
| 1992 | 0 | 43 | 10 | 18 | 3 | 0 | 0 | 0 | 74 |
| 1993 | 0 | 22 | 22 | 2 | 1 | 0 | 0 | 0 | 47 |
| 1994 | 4 | 8 | 19 | 12 | 0 | 0 | 0 | 0 | 43 |
| 1995 | 2 | 115 | 19 | 7 | 1 | 0 | 0 | 0 | 143 |
| 1996 | 0 | 28 | 40 | 7 | 1 | 0 | 0 | 0 | 77 |
| 1997 | 0 | 14 | 8 | 3 | 1 | 0 | 0 | 0 | 26 |
| 1998 | 2 | 7 | 4 | 6 | 3 | 0 | 0 | 0 | 23 |
| 1999 | na | na | na | na | na | na | na | na | na |
| 2000 | na | na | na | na | na | na | na | na | na |
| 2001 | na | na | na | na | na | na | na | na | na |
| 2002 | 31 | 207 | 72 | 21 | 9 | 1 | 0 | 0 | 340 |
| 2003 | 1 | 68 | 69 | 21 | 3 | 0 | 0 | 0 | 163 |
| 2004 | 32 | 28 | 29 | 9 | 5 | 0 |  | 0 | 102 |
| 2005 | 47 | 123 | 35 | 7 | 5 | 1 | 3 | 0 | 221 |
| 2006 | 32 | 148 | 60 | 24 | 1 | 1 | 0 | 0 | 170 |
| 2007 | 7 | 170 | 82 | 15 | 1 | 0 | 0 | 0 | 275 |
| 2008 | na | na | na | na | na | na | na | na | na |
| 2009 | na | na | na | na | na | na | na | na | na |
| 2010 | 138 | 155 | 120 | 58 | 12 | 1 | 0 | 0 | 484 |
| 2011 | 20 | 526 | 106 | 44 | 19 | 1 | 0 | 0 | 717 |
| 2012 | 7 | 184 | 304 | 30 | 8 | 3 | 0 | 0 | 536 |
| 2013 | 4 | 158 | 105 | 104 | 27 | 8 | 1 | 1 | 408 |
| 2014 | 7 | 46 | 45 | 25 | 19 | 4 | 0 | 1 | 146 |
| 2015 | 2 | 39 | 44 | 59 | 49 | 39 | 3 | 1 | 236 |

Table 15.3.1.2, continued : NAFO Div. 1D. Cod abundance indices (numbers of cod caught per 100 hours net settings) by age in the West Greenland inshore gillnet survey.

| Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ | All |
| 1985 | 68 | 77 | 0 | 3 | 3 | 3 | 0 | 1 | 155 |
| 1986 | 0 | 96 | 15 | 0 | 0 | 1 | 2 | 0 | 114 |
| 1987 | 1 | 16 | 68 | 5 | 0 | 0 | 0 | 0 | 90 |
| $1988$ | 0 | 20 | 48 | 30 | 1 | 0 | 0 | 0 | 99 |
| 1989 | 0 | 78 | 47 | 13 | 13 | 0 | 0 | 0 | 152 |
| $1990$ | 0 | 14 | 35 | 4 | 4 | 3 | 0 | 0 | 60 |
| $1991$ | 124 | 3 | 17 | 6 | 2 | 1 | 0 | 0 | 154 |
| 1992 | 0 | 61 | 22 | 10 | 7 | 1 | 0 | 0 | 100 |
| $1993$ | 0 | 4 | 57 | 20 | 2 | 0 | 0 | 0 | 83 |
| $1994$ | 0 | 0 | 6 | 5 | 1 | 0 | 0 | 0 | 12 |
| $1995$ | 0 | 3 | 2 | 4 | 4 | 0 | 0 | 0 | 12 |
| 1996 | 0 | 1 | 1 | 0 | 2 | 0 | 0 | 0 | 4 |
| 1997 | 3 | 3 | 1 | 0 | 0 | 1 | 0 | 0 | 8 |
| 1998 | 0 | 10 | 17 | 1 | 0 | 0 | 0 | 0 | 28 |
| 1999 | 0 | 0 | 1 | 3 | 0 | 0 | 0 | 0 | 5 |
| 2000 | 0 | 2 | 2 | 1 | 1 | 0 | 0 | 0 | 6 |
| $2001$ | na | na | na | na | na | na | na | na | na |
| 2002 | 0 | 7 | 4 | 3 | 0 | 0 | 0 | 0 | 14 |
| 2003 | 0 | 6 | 4 | 2 | 1 | 0 | 0 | 0 | 13 |
| 2004 | 3 | 43 | 6 | 3 | 1 | 1 | 0 | 0 | 57 |
| 2005 | 9 | 27 | 7 | 2 | 0 | 0 | 0 | 0 | 45 |
| 2006 | 2 | 114 | 37 | 13 | 4 | 0 | 0 | 0 | 170 |
| 2007 | na | na | na | na | na | na | na | na | na |
| 2008 | 4 | 4 | 47 | 63 | 7 | 0 | 0 | 0 | 124 |
| 2009 | 4 | 52 | 14 | 72 | 23 | 1 | 0 | 0 | 166 |
| 2010 | 1 | 33 | 107 | 18 | 27 | 3 | 0 | 0 | 189 |
| 2011 | 10 | 45 | 3 | 18 | 6 | 4 | 1 | 0 | 88 |
| 2012 | 2 | 52 | 46 | 21 | 28 | 2 | 0 | 1 | 151 |
| 2013 | 0 | 91 | 61 | 77 | 25 | 8 | 3 | 2 | 267 |
| 2014 | 0 | 41 | 74 | 46 | 27 | 6 | 1 | 0 | 196 |
| 2015 | 2 | 42 | 79 | 68 | 30 | 7 | 2 | 0 | 229 |

Table 15.3.1.2, continued : NAFO Div. 1F. Cod abundance indices (numbers of cod caught per 100 hours net settings) by age in the West Greenland inshore gillnet survey.

| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ | All |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 204 | 8 | 1 | 1 | 1 | 1 | 1 | 0 | 217 |
| 1986 | 17 | 112 | 5 | 0 | 2 | 0 | 0 | 0 | 136 |
| 1987 | 0 | 143 | 147 | 1 | 0 | 0 | 0 | 0 | 291 |
| 1988 | 0 | 1 | 83 | 6 | 0 | 0 | 0 | 0 | 89 |
| 1989 | 0 | 5 | 2 | 19 | 2 | 0 | 0 | 0 | 29 |
| 1990 | 0 | 0 | 3 | 2 | 13 | 1 | 0 | 0 | 18 |
| 1991 | 2 | 2 | 0 | 2 | 0 | 1 | 0 | 0 | 7 |
| 1992 | 0 | 3 | 1 | 0 | 1 | 0 | 1 | 0 | 6 |
| 1993 | 0 | 5 | 2 | 1 | 0 | 0 | 0 | 0 | 8 |
| 1994 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 3 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | na | na | na | na | na | na | na | na | na |
| 1997 | na | na | na | na | na | na | na | na | na |
| 1998 | 0 | 4 | 12 | 0 | 0 | 0 | 0 | 0 | 17 |
| 1999 | na | na | na | na | na | na | na | na | na |
| 2000 | 0 | 14 | 8 | 0 | 2 | 0 | 1 | 0 | 24 |
| 2001 | na | na | na | na | na | na | na | na | na |
| 2002 | na | na | na | na | na | na | na | na | na |
| 2003 | na | na | na | na | na | na | na | na | na |
| 2004 | na | na | na | na | na | na | na | na | na |
| 2005 | na | na | na | na | na | na | na | na | na |
| 2006 | na | na | na | na | na | na | na | na | na |
| 2007 | 6 | 90 | 9 | 21 | 1 | 0 | 0 | 0 | 108 |
| 2008 | 8 | 17 | 30 | 4 | 2 | 0 | 0 | 0 | 62 |
| 2009 | 3 | 39 | 14 | 15 | 0 | 0 | 0 | 0 | 71 |
| 2010-2015 | na | na | na | na | na | na | Na | na | na |

Table 15.5.1. Number of tagged cod in the period of 2003 to 2014 in different regions. Bank (West) = NAFO division 1D+1E. East Greenland = NAFO division 1F + ICES division 14.b.

| Tagged |  |  |  |
| :---: | :---: | :---: | :---: |
| Year | FJord | Bank (West) | East Greenland |
| 2003 | 599 |  |  |
| 2004 | 658 |  |  |
| 2005 | 565 |  | 1387 |
| 2006 | 41 |  | 1296 |
| 2007 | 1140 |  | 525 |
| 2008 | 231 |  |  |
| 2009 | 633 | 1563 | 403 |
| 2010 | 88 | 2321 | 2359 |
| 2011 | 28 |  | 1203 |
| 2012 | 86 | 57 | 1218 |
| 2013 | 183 |  |  |
| 2014 |  |  |  |
| 2015 |  |  |  |

Table 15.5.2: Number of recaptured cod in the period of 2003 to 2014 in different regions. Fjord (West) = NAFO divisions 1B-1F. Bank (West) = NAFO division 1D+1E. East Greenland = NAFO division 1F + ICES division 14.bb.

| Recaptures |  |  |  |
| :--- | :---: | :---: | :---: |
|  | FJord (West) | Bank (West) | East Greenland |
| Fjord (West) | 438 | 14 | 2 |
| Bank (West) | 1 | 34 | 2 |
| East Greenland |  | 12 | 99 |
| Iceland | 3 | 21 | 125 |

### 15.13Figures



Figure 15.2.1.1 Inshore landings from West Greenland (Horsted 1994, 2000).


Figure 15.2.1.3. Total catches and TAC in the inshore fishery by NAFO Divisions.


Figure 15.2.1.2. Distribution of commercial fishery along the coastline of West Greenland in total tons by field code.


Figure 15.2.1.3. Distribution of the inshore commercial fishery by gear (tons/fieldcode).


Figure 15.2.2.1. Total length and age distributions of inshore cod catches.

## Inshore CAA commercial fishery



Figure 15.2.2.2. Catch-at-age in the commercial fishery in the West Greenland inshore area. Size of circles represents size of catch numbers.


Figure 15.3.1.1. The inshore gillnet survey area on the Greenland West coast. Top picture is the Sisimiut fjord system in NAFO 1B and bottom picture is the Nuuk fjord system in NAFO 1D. Survey estimates of catch rates are indicated on both maps as \#caught/100h.


Figure 15.3.1.2. : Recruitment indices (numbers caught/100 hr.) for ages 2 and 3 in 1B (top), 1D (middle) and 1B and 1D combined (lower) in West Greenland. Simultaneous surveys were not carried out 1999-2001 and 2007-2009.


Figure 15.3.1.3. Recruitment indices (numbers caught/100 hr.) for ages 1-5 in 1B (left), 1D (middle) and 1B and 1D combined (right) in West Greenland from 1985-2015. Size of circles represents the size of the index values and the values are standardized within each area and are not comparable among each other.


Figure 15.3.1.4. Abundance indices (numbers caught/ 100 hrs . netsetting) for all age groups (left) and age 3-8 (right) in 1B and 1D combined. Simultaneous surveys were not carried out 1999-2001 and 2007-2009.


Figure 15.10.1. Sum of $2+3$ yr old against sum of $4+5$ year old two years later in the gillnet survey.


Figure 15.10.2 Internal consistency plot of the gillnet survey.

## 16 Cod in offshore waters of ICES Subarea 14 and NAFO subarea 1

### 16.1 Stock definition

The cod found in Greenland is derived from four separate "stocks" that each is labelled by their spawning areas: I) offshore West Greenland waters; II) West Greenland fiords; III) offshore East Greenland and Icelandic waters and IV) inshore Icelandic waters (Therkildsen et al. 2013), (Figure 16.1).

From 2012 the inshore component (West Greenland, NAFO Subarea 1) was assessed separately from all offshore components. From 2016 the offshore West Greenland (NAFO subdivisions 1A-E) and East Greenland (NAFO subdivision 1F and ICES Subarea 14) components was assessed separately. The Stock Annex provides more details on the stock identities including the references to primary works.

### 16.2 Fishery

### 16.2.1 The emergence and collapse of the Greenland offshore cod fisheries

The Greenland commercial cod fishery in East Greenland started in 1954, but started earlier in Southwest Greenland (NAFO subdivision 1F, table 16.2.2.1). The fishery gradually developed culminating with catch levels above 40,000 tons annually in the 1960s. Due to overfishing and deteriorating environmental conditions the stock size declined and the fishery completely collapsed in the early 1990's (Fig 16.2.1). In the 2000s catches have gradually increased with maximum catches in 2008 of 14500 tons. Between 2008 and 2010 offshore areal closures were implemented in order to protect the spawning stock in offshore areas. More details on the historical development in the fisheries are provided in the stock annex.

### 16.2.2 The offshore fishery in 2015

In 2014 a management plan was implemented for the offshore cod fishery in South and East Greenland. According to the management plan the TAC is 10000 tons/year between 2014 and 2016, The TAC between 2014 and 2016 is to be taken in equal amounts in four areas: Survey area Q1+Q2, Survey area Q3+Q4, Survey area Q5+Q6 and NAFO subdivision 1F. Fishery is not allowed during spawning season $1^{\text {st }}$ of April-31 ${ }^{\text {st }}$ of May. However a dispensation was given in 2014 and 2015 to fish in South Greenland (NAFO $1 \mathrm{~F}+\mathrm{Q} 5 \mathrm{Q} 6$ ) during these months under the assumption that the spawning stock is concentrated further to the north in East Greenland (corresponding to management area Q1Q2 and Q3Q4).

The management plan contains a TAC regulation rule that allows for the final TAC to be adjusted if survey results show an increase of more than $30 \%$ or a decrease of more than $15 \%$ compared to a reference period (2010-2013). The survey results for 2014 from both the Greenlandic and German survey showed an increase of more than $30 \%$ and the final TAC was set at 18,104 tons divided between management areas: 4,800t in Q1Q2, 4,000t in Q3Q4, 3,000t in Q5Q6, 3,949t in NAFO 1F. The rest of the TAC was given as $2,000 \mathrm{t}$ as bycatch in the whole area, and 355 t inshore in East Greenland. The TAC was divided between the following countries: 14,849 tons to Greenland, 2,000 tons to EU, 900 tons to Norway and 355 tons to the Faroe Islands.
In 2015 EU, Norway and the Faroe Islands fished their quota (Norway fished 330 t more than their quota) whereas Greenland fished 12,170 tons resulting in a total of

15,755 tons with 3,984 tons caught in SouthWest Greenland (NAFO 1F) and 11,771 tons caught in East Greenland (table 16.2.2.1).
$75 \%$ of the total catches were taken in East Greenland where the fishery took place throughout the year except in spawning time (April and May) where directed fishery for cod in Q1Q2 and Q3Q4 were not allowed. Catches in these areas in this period are considered as bycatch. Catches in SouthWest Greenland peaked in June and July (table 16.2.2.2). The fishery where distributed from Julianehåbs Bight in SouthWest Greenland $\left(60^{\circ} \mathrm{N}\right)$ to Dohrn Bank $\left(66^{\circ} \mathrm{N}\right)$ in East Greenland (figure 16.2.2.1).
A dispensation was given to Greenlandic longliners to fish within the 3 nm from the baseline in East Greenland where vessels larger than 75BRT/120BT are not allowed. In all the vessel caught almost 500 tons concentrated around the Tasiilaq area $\left(65-66^{\circ} \mathrm{N}\right)$ and Cape Fraewell (figure 16.2.2.1). No length measurements were available from this fishery.

Trawlers caught $67 \%$ of the total catch (table 16.2.2.3). Half of the total trawl catches were taken on Dohrn Bank (management area Q1Q2) especially in a small area between $65-66^{\circ} \mathrm{N}$ and $29-31^{\circ} \mathrm{W}$ on the edge of the continental shelf on Dohrn Bank (figure 16.2.2.2).

Longliners caught $33 \%$ of the total catch (table 16.2.2.3). The longliners fished primarily north of Skjoldungen Bank and around Kleine Bank in management area Q3Q4 and in Julianehåbs Bight in Southwest Greenland (NAFO 1F, figure 16.2.2.2).

### 16.2.3 Length, weight and age distributions in the offshore fishery 2015

There is limited landing sample information from the 1990's where the cod fishery was very low in East Greenland. For that period length frequency information is generally lacking for the offshore fisheries where cod was only taken as a bycatch. Sampling intensities have increased considerably in the later years, and in 2015 the offshore fisheries was very well covered.

Catch-at-age and weight-at-age has been compiled for the offshore area since 2005 (table 16.2.3.1).

Length measurement in 2015 in SouthWest Greenland amounted to 7144 cod measured. In East Greenland length measurements amounted to 4900.

The overall mean length in the catches was 70 cm , and the YC 2009 (6 yr old fish) dominated the catches (figure 16.2.3.1). Mean length differed however between areas with the largest cod (mean length $=78 \mathrm{~cm}$ ) and oldest ( $7-8 \mathrm{yr}$ old) being caught on Dohrn Bank furthest to the north in the fishing area in East Greenland (management area Q1Q2), and the smallest (mean length $=62 \mathrm{~cm}$ ) and youngest ( 6 yr old) being caught in SouthWest Greenland (NAFO 1F, figure 16.2.3.2). Especially in Southwest Greenland the 2009 YC ( 6 yr old) dominated the catches with $50 \%$ of the catch being of this YC.
In 2012, 2013 and 2014 the 2007 YC dominated the total catches as 5,6 and 7 yr olds (Table 16.2.3.1, figure 16.2.3.3). This YC was especially abundant in the catches in South Greenland in 2013. In 2014 this YC was abundant in all areas but in 2015 the YC is more abundant in the northern part of the fishing area in East Greenland (management area Q1Q2, figure 16.2.3.2). The 2008 YC and 2009 YC were in 2014 (as 6 and 5 yr old) more abundant in SouthWest Greenland (NAFO 1F). In 2015 the 2008 YC is dominating in East Greenland as 7 yr old whereas the 2009 YC is dominating in Southwest Greenland as 6 yr old (figure 16.2.3.2).

### 16.2.4 cpue index

Logbooks on a haul by haul basis from the cod fishery since 1975 where compiled in 2014. But due very low catches and few hauls in the 90 'ies and closed areas in 20082010, the logbook data are not used in the assessment process. Nevertheless, cpue results generated by a GLM model are presented here.
As EU and Greenland vessels have participated in the fisheries in the entire period, data from these were used in the GLM model. Hauls made in the closed area in the period of 2008-2010 were excluded from the analysis, as they were considered being bycatches.
The cpue index was relatively stable in the first part of the time-series (1975-1992, mean 0.673 ton $/ \mathrm{h}$ ), except 1989 where cpue increased to 1.668 ton/h (table 16.2.4.1, figure 16.2.4.1). This increase was likely caused by the large 1984 YC entering the fishery. cpue then declined from 1993-2005 ( 0.130 ton/h) but sampling of the fishery was low in this period due to very low catches of about 200-300 tons total, and catches where taken primarily as bycatch in the redfish fishery. In 2006-2008 cpue increased (mean 1.632 ton $/ \mathrm{h}$ ) as catches started to increase. In 2009 however cpue decreased to 0.397 ton/h, which was most likely caused by the east ward migration of the 2003 YearClass out of the allowed fishing areas.

In 2010, where almost all of the offshore area was closed except of a small area in Southeast Greenland, the index increased to 0.655 ton $/ \mathrm{h}$, but catches were taken by very few vessels. In 2011 all closed areas where reopened and fishery started again especially in East Greenland north of $63^{\circ} \mathrm{N}$ resulting in an increase in cpue to 1.206 ton/h. Since 2011 cpue has declined to 0.420 ton/h in 2015 where cpue where half of the cpue in 2014.

Since 2011 the cpue index has declined and was in 2015 half of the cpue in 2014. In contrast the survey biomass index increased from 2012 to 2013 which is properly caused by the large 2009 YC entering the area and observed in the survey. This Yearclass however was not observed in the fishery until 2015 as the fishery in 2013 and 2014 where concentrated more to the north in East Greenland and therefore fished older and larger fish. The downwards trend in cpue index since 2013/2014 is also observed in the survey biomass index.

### 16.3 Surveys

At present, two offshore trawl surveys (Greenlandic and German) provide the core information relevant to stock assessment purposes. For details of survey design see stock annex.
The German survey targets mainly cod and has since 1982 covered the main cod grounds off both East and West Greenland at depths down to 400 m . The Greenland survey in West Greenland targets shrimp and cod down to 600 m . The Greenland survey is believed to provide a better coverage of the cod distribution in especially East Greenland as the survey has twice as many stations covering both shelf edge and top, whereas the stations in the German survey are usually concentrated at the shelf edge. The Greenland survey time-series is however limited as the survey in East Greenland first started in 2008.

### 16.3.1 Results of the Greenland Shrimp and Fish survey in South and East Greenland

A total number of 131 valid hauls were made in 2015 (Error! Reference source not found.).

For Atlantic cod the abundance index was estimated at 63 million individuals and the survey biomass at 154,700 tons. Survey abundance increased slightly with $9 \%$ compared to 2014, whereas biomass decreased with $16 \%$. Half of the total abundance ( $53 \%$ ) and one third of the total biomass ( $32 \%$ ) was found in SouthWest Greenland (NAFO 1F) (Error! Reference source not found.16.3.1.2, 16.3.1.3, Error! Reference source not found.16.3.1.1 and 16.3.1.2). 21\% of the total biomass was found on Dohrn Bank (survey area Q1), whereas only $10 \%$ of the total abundance was found here indicating that large cod inhabits this area.

The dominating cohort is the 2009 YC accounting for $46 \%$ in abundance and the second largest cohort was the 2010 YC (18\%) (Table 16.3.1.4, figure 16.3.1.3). The 2009 YC is dominating in South Greenland (NAFO 1F and Q6) where $84 \%$ of the total 2009 YC abundance is found (table 16.3.1.5, figure 16.3.1.4). The 2010 YC is more dominant in Southwest Greenland (NAFO 1F) where $80 \%$ of the total 2010 YC abundance is found. In 2014 the 2007 YC was the second largest YC in the survey. In 2015 this YC is the highest registered number of 8-yr old in the time-series (since 2008) and as in 2014 this YC is distributed further to the north in East Greenland (Dohrn Bank (Q1) - Kleine Bank (Q3)).

In general younger cod (3-6 yrs) are predominantly found in South Greenland (NAFO $1 \mathrm{~F}+\mathrm{Q} 6$ ), whereas older cod (> 7 yrs ) are found in the northern survey area in East Greenland (table 16.3.1.5, figure 16.3.1.5).
SSB index was estimated to 150,000 t in 2014 which was the highest observed in the time-series. In 2015 index is estimated to 119,000t which is a decrease of $20 \%$ (figure 16.3.1.7). The spawning stock is distributed throughout the area, but is highest in Southwest (1F), Mideast (Q3) and Northeast (Q1) areas (figure 16.3.1.8). Spawning has however not been documented in Southwest Greenland as the survey is conducted outside the spawning season.

The smoothed biomass estimates are very close to the observed mean estimates in 2015 (table 16.3.1.3, figure 16.3.1.9) where the CV was low. The observed CV was 0.20 , while the smoothed CV estimate was 0.18 . The process SD was 0.32 .

### 16.3.2 Results of the German groundfish survey off West and East Greenland

In 2015, 75 valid trawl stations were sampled during autumn in the German Greenland offshore groundfish survey (Table 16.3.2.1, Figure 16.3.2.1).

Abundance and biomass decreased by 95\% and 88\% respectively from 2014-2015 (table 16.3.2.2 and 16.3.2.3). The indices in 2015, both abundance and especially biomass, was highest recorded in the time-series since 1982 (figure 16.3.2.4 and 16.3.2.5). In contrast the indices in 2015 was the lowest recorded since 2011. The main reason for the increase in abundance and biomass since 2012 was increased numbers of the 2009 YC, which has dominated the survey in the period 2012-2014 (Table 16.3.2.4). The large increase in biomass is probably caused by a southward migration of the 2009 YC, which at a younger age was observed more in West Greenland than in South Greenland. In 2015 the 2009 YC was not caught in considerable numbers and it was especially missing in Southwest Greenland were the largest decline in abundance and biomass was observed. The 2010 YC first appeared in considerable numbers in the survey in 2014
where it was the second largest YC observed. In 2015 the 2009 and 2010 YC was dominating the survey (at age 6 and 5) but with a decrease of $94 \%$ (2009 YC) and $82 \%$ (2010 YC) compared with their indices at age 4 and 5 in 2014 (table 16.3.2.4).

The survey time-series (figure 16.3.2.4) shows three abundance peaks in 1987-1989 caused by the 1984 and 1985 YC, in 2005-2007 caused by the 2003 YC and in 20132014 caused by the 2009 YC. Biomass indices show the same peaks, although a large increase in biomass in 2014 compared to the previous periods (figure 16.3.2.5).

Overall findings where the same between the Greenland and the German survey, although the German survey show large fluctuations: a 2009 YC dominating the catches in recent years in South Greenland. However the German survey observed dramatic declining numbers of the 2009 YC in especially Southwest Greenland that was not observed in the Greenland survey. The German survey is conducted in October, whereas the Greenland survey in West Greenland is conducted in June/July. The Greenland survey has twice as many stations and a wider coverage area.

Both surveys show that older and larger cod are found furthest to the north in East Greenland, especially in the Dohrn Bank region.

### 16.4 Information on spawning

Adequate maturity information has been lacking for the offshore cod stock as the Greenland and German surveys are conducted well outside the spawning period. The offshore fishery has however shown dense concentrations of large spawning cod off East Greenland at least since 2004. The fishery showed that spawning is concentrated on-banks north of $62^{\circ} \mathrm{N}$ in East Greenland. For further information on spawning see stock annex.

### 16.5 Tagging experiments

A total of 17304 cod have been tagged in different regions of Greenland in the period of 2003-2015 (table 16.5.1). Cod in the offshore area in West Greenland have been tagged in 2007, 2012 and 2013 on Dana Bak (NAFO 1DE). Cod offshore in East Greenland have been tagged in 2007-2009, 2011, 2012 and 2014-2015 from Julianhåbs Bight (NAFO 1F) in SouthWest Greenland to Dohrn Bank in East Greenland.

Inshore recaptures are almost exclusively recaptured in the same place as tagged (table 16.5.2). No tags from the inshore area have been recaptured offshore except 3 that were recaptured in Iceland. These three cod were tagged in the inshore area in South Greenland.

Offshore recaptures are found both in West-, East Greenland and Iceland (table 16.5.2). Most recaptured tags in both West Greenland are recaptured in the same place as they were tagged, but more tags tagged in East Greenland are recaptured in Iceland than in East Greenland (125 in Iceland compared to 99 in East Greenland). Fishing effort can influence the numbers of recaptures and more analysis needs to be performed on the tagging data in order to investigate the interaction between Iceland and East Greenland.

### 16.6 State of the stock

The offshore component has been severely depleted since 1990. However, the surveys indicate an improvement in recruitment with all year classes since 2002, and estimated at sizes above the very small year classes seen in the 1990s. These YC's has led to a stock increase during the 00 s and an increase in catches.

The overall trend in the two surveys is the same: the 2009 YC is distributed in South Greenland, whereas older yearclasses are distributed further north in East Greenland.

The German survey showed a doubling in biomass in 2014 and a reduction of $88 \%$ in 2015. The increase in 2014 was caused by increasing numbers of especially the 2009 YC, but also the 2010 YC in South Greenland. In 2015 the 2009 YC was not caught in significant numbers which caused the sharp decline in the survey.

The same increase in 2014 and sharp decline in 2015 was not observed in the Greenland survey and the biomass index in the survey seems stable over the past 3 years. The Greenland survey takes place during summer whereas the German survey takes place a couple of month later in autumn. Difference in season, haul numbers and coverage between the two surveys might explain the difference between the two surveys.

The fishery confirmed the distribution found in the surveys with younger yearclasses ( $<7 \mathrm{yrs}$ ) dominating the catches in South Greenland, and older yearclasses dominating the catches further north in East Greenland, especially in the Dohrn Bank area.

Indicators show that fishing pressure has been low the last $5-6$ yrs and the stock is considered to be improving. The stock size is however still low compared with the 1950's and 1960', where catches exceeded 30000 tons for a number of years.

### 16.7 Implemented management measures for 2016

The offshore quota for the total international fishery is set at 16000 tons divided into 4 management areas (see figure below).

To protect the spawning stock no fishing is allowed from April $1^{\text {st }}-$ May $31^{\text {st }}$ in all areas.

### 16.8 Management plan

In 2014 a management plan was implemented for the offshore cod fishery in Greenland (2014-2016) but it has not been evaluated by ICES. The management plan is built on the distinction between the inshore and two offshore stocks components.


Management area West covers NAFO Subarea 1A-E and management area Southeast covers ICES Subarea 14.b (survey area Q1-6) + NAFO 1F.

According to the management plan the TAC in management area Southeast is 10000 t/year between 2014 and 2016 and no fishery should be done north of 1F in West Greenland. However deviations have been the case in 2015 and 2016 where TAC was set higher than the proposed in the management plan.

The TAC in management area South and East Greenland is divided equally between four areas: Survey area Q1+Q2, Q3+Q4, Q5+Q6 and NAFO area 1F.

The management plan has not been evaluated by ICES.

### 16.9 Management considerations.

Larger and older fish ( $7+\mathrm{yr}$ old) are located furthest to the north on Dohrn Bank, whereas younger fish dominate in the South ( $5-7 \mathrm{yr}$ old). This reflects the eastward migration behaviour towards the spawning grounds in East Greenland. Further, the genetic studies combined with tagging results suggest that the spawning stock component in East Greenland is associated with the offshore spawning population in Iceland, but the extent and exact dynamics of this association is unknown.

### 16.10 Basis for advice

This stock was benchmarked in 2015 and the benchmark concluded that catch advice for this stock should be based on the 3.2 DLS approach (ICES 2012, ICES 2015). The NWWG however concluded that this approach had some major drawbacks for this stock. At NWWG the applicability of the DLS approach was explored and several shortcomings in relation to this stock were found:

- Using the DLS is a slow responding approach not suited to a species with a very dynamic stock development. Applying it for this stock would not allow managers to react to sudden increases or decreases in biomass due to the $20 \%$ cap/change limit from catches. To adjust for this, additional exploratory
analyses looked at the consequence of having a cap $10 \%, 20 \%, 40 \%$ and without (Figure 16.10.1). The $40 \%$ and no cap however entailed that the advice rose very quickly in response to increasing survey values. On the other hand, the lower caps were not able to react in periods with low biomasses.
- The level of advised catch depends to a very large extent on the offset (Figure 16.10.2). For instance, if the approach is implemented at a time of low catches $(<500 \mathrm{t})$ it will take a long time for the advice to adjust to increasing biomass. The other case with a starting point with high catches could also be chosen and that would result in an advice that starts high and stays at that level for a long time.

These issues raised above are particularly important for this stock were large year-toyear variations are a natural occurrence. Therefore, the NWWG concluded that the DLS category 3.2 method was not the best available option, and instead DLS category 3.3 (ICES 2012) with an $\mathrm{F}_{\text {proxy }}$ as a reference point was a better alternative. As a period of relatively stable catches is co-occurring with rising survey indices (2011-2014), a derived $\mathrm{F}_{\text {proxy }}$ would be a better basis for advice and more precautionary. The fishing mortality in this period was explored by log catch ratios (Figure 16.10.3) and NWWG concluded that as F appeared to be very low in this period no precautionary buffer should be applied. Also, as the stock status is well described through two surveys no uncertainty cap should be applied. Hence, the catch advice should be based on an $\mathrm{F}_{\text {proxy }}$ multiplier on the Greenland survey (smoothed) which has the best coverage of the stock. The catch was divided by the survey from 2011-2014 and the average of this (0.049) was multiplied with the smoothed 2015 Greenland survey index (157 312) to give the 2017 catch advice.

### 16.11References

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### 16.12Tables

Table 16.2.2.1. Offshore catches $(\mathbf{t})$ divided into NAFO divisions in West Greenland and East Greenland (ICES 14.b). 1924-1995: Horsted 2000, 1995-2000: ICES Catch Statistics, 2001-present: Greenland Fisheries License Control.



|  | NAFO |  |  |  |  |  | UnKNOWN |  | NAFO |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  | $1 \mathrm{~F}+$ |
|  |  | NAFO | NAFO | NAFO | NAFO | NAFO |  | ICES | ICES |
| Year | 1 A | 1 B | 1 C | 1D | 1E | 1 F | NAFO DIV. | $14 . \mathrm{B}$ | $14 . \mathrm{B}$ |
| 2005 | 0 | 0 | 1 | 0 | 0 | 71 | 0 | 819 | 890 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 414 | 0 | 2042 | 2456 |
| 2007 | 0 | 0 | 0 | 31 | 435 | 20113 | 0 | 3194 | 5205 |
| 2008 | 0 | 0 | 0 | 23 | 526 | 113703 | 0 | 3258 | 14628 |
| 2009 | 0 | 0 | 0 | 0 | 6 | 33233 | 0 | 1642 | 4965 |
| 2010 | 0 | 0 | 0 | 0 | 2 | 281 | 0 | 2388 | 2669 |
| 2011 | 0 | 0 | 0 | 0 | 8 | 542 | 0 | 4571 | 5113 |
| 2012 | 0 | 0 | 1 | 95 | 236 | 1470 | 0 | 3941 | 5411 |
| 2013 | 0 | 0 | 0 | 209 | 270 | 1405 | 0 | 4104 | 5509 |
| 2014 | 0 | 0 | 30 | 68 | 18 | 1833 | 0 | 6060 | 7893 |
| 2015 | 0 | 0 | 341 | 954 | 3564 | 3984 | 0 | 11771 | 15755 |

Estimates for assessment include estimates of unreported catches. The total estimated value for West Greenland (inshore + offshore) was 73000 t in 1977 and 1978, 1979: 99000 t , 1980: 54000 t . The value given in the table are these values minus the inshore catches minus known offshore NAFO division catches.
Estimates for assessment include estimates of unreported catches in East Greenland.
Include catches taken with small vessels and landed to a factory in South Greenland (Qaqortoq), 2007:597 t, 2008: 2262 t, 2009: 136 t.
*) Unknown NAFO division catches added accordingly to the proportion of known catch in NAFO division 1F to known total catch in all NAFO divisions.

Table 16.2.2.2: 2015 cod catches (t) by area and month. East Greenland (14.b) divided into three management areas.

| ICES/NAF <br> O | JAN | FEB | MAR | R | MAP | JUN | JUL | AUG | SEP | T | NOV | DEC | L | \% |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 14.b |  |  |  |  |  |  |  |  |  |  |  | 137 |  | 34 |
| (Q1Q2) | 423 | 88 | 509 | 78 | 32 | 689 | 585 | 118 | 691 | 430 | 392 | 8 | 5413 | $\%$ |
| 14.b | 138 | 15 |  |  |  |  |  |  |  |  |  |  |  | 24 |
| (Q3Q4) | 9 | 3 | 77 | 15 | 127 | 613 | 576 | 478 | 185 | 4 | 0 | 212 | 3829 | $\%$ |
| 14.b |  |  |  |  |  |  |  |  |  |  |  |  |  | 16 |
| (Q5Q6) | 83 | 6 | 219 | 319 | 660 | 75 | 41 | 122 | 201 | 297 | 244 | 262 | 2529 | $\%$ |
| 1F |  | 23 |  |  |  |  |  |  |  |  |  |  |  | 25 |
|  | 265 | 3 | 226 | 160 | 334 | 701 | 786 | 510 | 117 | 9 | 448 | 196 | 3984 | $\%$ |
| Total | 216 | 47 | 103 |  | 115 | 207 | 198 | 122 | 119 |  | 108 | 204 |  |  |
| $\%$ | 0 | 9 | 0 | 572 | 3 | 8 | 8 | 7 | 4 | 740 | 5 | 9 | 15755 |  |
| $\%$ | $14 \%$ | $3 \%$ | $7 \%$ | $4 \%$ | $7 \%$ | $13 \%$ | $13 \%$ | $8 \%$ | $8 \%$ | $5 \%$ | $7 \%$ | $13 \%$ |  |  |

Table 16.2.2.3: 2015 cod catches $(t)$ by gear, area and month. East Greenland (14.b) divided into three management areas.

| Gear | ICES/NAFO | JAN | Feb | Mar | APR | May | JuN | JUL | Aug | SEP | Oct | Nov | Dec | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Longline | $\begin{aligned} & \text { 14.b } \\ & \text { (Q1Q2) } \end{aligned}$ |  |  |  |  |  |  | 13 | 65 | 51 | 3 |  |  | 131 |
|  | $\begin{aligned} & 14 . \mathrm{b} \\ & \text { (Q3Q4) } \end{aligned}$ | 958 |  |  |  | 95 | 249 | 26 | 328 | 168 | 0.02 |  |  | 1824 |
|  | $\begin{aligned} & 14 . \mathrm{b} \\ & \text { (Q5Q6) } \end{aligned}$ |  |  |  | 45 | 29 |  |  | 120 | 197 | 296 | 242 | 26 | 955 |
|  | 1F | 265 | 233 | 226 | 160 | 233 | 377 | 233 | 141 | 19 | 9 | 168 | 175 | 2237 |
|  | Total | 1223 | 233 | 226 | 205 | 357 | 626 | 272 | 654 | 435 | 308 | 410 | 201 | 5148 |
| Trawl | $\begin{aligned} & 14 . b \\ & \text { (Q1Q2) } \end{aligned}$ | 423 | 88 | 509 | 78 | 32 | 689 | 572 | 53 | 640 | 427 | 392 | 1378 | 5282 |
|  | $\begin{aligned} & 14 . \mathrm{b} \\ & \text { (Q3Q4) } \end{aligned}$ | 431 | 153 | 77 | 15 | 32 | 364 | 550 | 150 | 16 | 4 |  | 212 | 2004 |
|  | $\begin{aligned} & 14 . \mathrm{b} \\ & \text { (Q5Q6) } \end{aligned}$ | 83 | 6 | 219 | 273 | 632 | 75 | 41 | 2 | 4 | 1 | 2 | 236 | 1574 |
|  | 1F |  |  |  |  | 102 | 324 | 553 | 369 | 98 |  | 280 | 21 | 1747 |
|  | Total | 937 | 247 | 804 | 367 | 797 | 1452 | 1716 | 574 | 759 | 432 | 675 | 1848 | 10607 |

Table 16.2.3.1. Cod in Greenland. Catch-at-age ('000) and Weight at age ( $\mathbf{k g}$ ) for offshore fleets in East Greenland (ICES 14.b + NAFO 1F).

| CATCH-AT-AGE |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year/age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 2005 | 5 | 33 | 57 | 103 | 94 | 57 | 16 | 7 |
| 2006 | 232 | 376 | 135 | 175 | 115 | 14 | 1 | 0 |
| 2007 | 49 | 1529 | 668 | 158 | 124 | 120 | 18 | 15 |
| 2008 | 77 | 586 | 6015 | 2417 | 592 | 44 | 26 | 12 |
| $2009$ | 307 | 1287 | $1231$ | 434 | 119 | 28 | 16 | 2 |
| $2010$ | 10 | 87 | 331 | 193 | 334 | 58 | 8 | 5 |
| $2011$ | 3 | 70 | 137 | 425 | 355 | 371 | 96 | 31 |
| 2012 | 13 | 109 | 471 | 281 | 258 | 253 | 148 | 59 |
| $2013$ | 0 | 36 | 127 | 615 | 237 | 226 | 153 | 104 |
| 2014 | 1 | 4 | 279 | 434 | 658 | 335 | 173 | 131 |
| 2015 | 3 | 57 | 457 | 1554 | 1324 | 828 | 242 | 182 |
| Weight at age |  |  |  |  |  |  |  |  |
| 2005 | 0.354 | 0.717 | 1.073 | 1.963 | 2.737 | 3.699 | 5.271 | 7.366 |
| 2006 | 1.323 | 1.602 | 2.349 | 3.608 | 4.420 | 5.440 | 7.191 | 8.127 |
| $2007$ | $0.387$ | 0.917 | 1.597 | 3.294 | 6.092 | 8.524 | 11.114 | 14.435 |
| 2008 | $0.359$ | 0.644 | 1.266 | 1.799 | 3.025 | 4.936 | 5.840 | 8.290 |
| 2009 | 0.489 | 0.776 | 1.396 | 2.797 | 4.634 | 6.453 | 7.804 | 9.993 |
| 2010 | 0.699 | 1.125 | 1.636 | 2.494 | 3.354 | 5.334 | 8.063 | 10.475 |
| 2011 | 0.553 | 1.026 | 1.541 | 2.297 | 3.377 | 4.685 | 6.285 | 10.022 |
| 2012 | 0.502 | 0.892 | 1.440 | 2.380 | 3.570 | 5.142 | 7.172 | 11.417 |
| 2013 | 0.480 | 0.998 | 1.698 | 2.272 | 3.408 | 4.745 | 6.827 | 9.024 |
| 2014 | 0.564 | 1.163 | 1.853 | 2.603 | 3.636 | 4.732 | 6.400 | 8.841 |
| 2015 | 0.484 | 0.833 | 1.435 | 2.097 | 3.460 | 4.699 | 6.846 | 9.115 |

Table 16.2.4.1: Data used in the Atlantic cod cpue. N are number of hauls from vessels from EU and Greenland used in the analysis.

| year | N | LN CPUE (ton/h) | SE |
| :---: | :---: | :---: | :---: |
| 1975 | 82 | -1.1356837 | 0.1837926 |
| 1976 | 5 | -0.9648581 | 0.6916295 |
| 1977 | 304 | 0.0759557 | 0.1199237 |
| 1978 | 232 | -0.2357506 | 0.1298423 |
| 1979 | 313 | -0.1948766 | 0.1294628 |
| 1980 | 106 | -0.8420723 | 0.1677385 |
| 1981 | 10 | -1.4465793 | 0.4936838 |
| 1982 | 15 | -1.2515564 | 0.4022425 |
| 1983 | 52 | -0.7002941 | 0.2559437 |
| 1984 | 211 | -0.5794715 | 0.1554628 |
| 1985 | 41 | -0.3401624 | 0.2552333 |
| 1986 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 0 |
| 1988 | 368 | -0.0616769 | 0.0998585 |
| 1989 | 1637 | 0.5118464 | 0.0753617 |
| 1990 | 4374 | -0.0431842 | 0.0492198 |
| 1991 | 3007 | -0.6929048 | 0.0524628 |
| 1992 | 2392 | -0.5291326 | 0.0563145 |
| 1993 | 244 | -2.1702938 | 0.1122777 |
| 1994 | 124 | -3.7735774 | 0.1456914 |
| 1995 | 6 | -3.7388718 | 0.627408 |
| 1996 | 123 | -2.2475193 | 0.19314 |
| 1997 | 16 | -0.8358301 | 0.3881681 |
| 1998 | 40 | -2.4133971 | 0.2625715 |
| 1999 | 177 | -2.5887944 | 0.1651539 |
| 2000 | 22 | -2.2269905 | 0.3293105 |
| 2001 | 94 | -2.0320493 | 0.1657239 |
| 2002 | 140 | -3.0407629 | 0.1863067 |
| 2003 | 144 | -1.8944053 | 0.1504092 |
| 2004 | 89 | -2.3750714 | 0.2123546 |
| 2005 | 55 | -1.2238767 | 0.4065565 |
| 2006 | 261 | 0.4300269 | 0.1222363 |
| 2007 | 358 | 0.7340622 | 0.0913542 |
| 2008 | 1530 | 0.2447458 | 0.0636558 |
| 2009 | 710 | -0.9229665 | 0.0800496 |
| 2010 | 255 | -0.4225995 | 0.1151019 |
| 2011 | 500 | 0.1876315 | 0.0898746 |
| 2012 | 493 | -0.0729454 | 0.0893399 |
| 2013 | 435 | -0.3109074 | 0.1062124 |
| 2014 | 947 | -0.0703275 | 0.0892882 |
| 2015 | 1814 | -0.867223 | 0.0772261 |
| Total | 21726 |  |  |

Table 16.3.1.1. Number of hauls in the Greenland Shrimp and Fish survey in ICES 14.b and NAFO 1F.

| Year/Strata | Q1 | Q2 | ICES $14 . \mathrm{B}$ |  | NAFO |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Q3 | Q4 | Q5 | Q6 | 1 F | Total |
| 1992 |  |  |  |  |  |  | 15 |  |
| 1993 |  |  |  |  |  |  | 13 |  |
| 1994 |  |  |  |  |  |  | 9 |  |
| 1995 |  |  |  |  |  |  | 11 |  |
| 1996 |  |  |  |  |  |  | 11 |  |
| 1997 |  |  |  |  |  |  | 19 |  |
| 1998 |  |  |  |  |  |  | 14 |  |
| 1999 |  |  |  |  |  |  | 17 |  |
| 2000 |  |  |  |  |  |  | 29 |  |
| 2001 |  |  |  |  |  |  | 26 |  |
| 2002 |  |  |  |  |  |  | 27 |  |
| 2003 |  |  |  |  |  |  | 22 |  |
| 2004 |  |  |  |  |  |  | 34 |  |
| 2005 |  |  |  |  |  |  | 23 |  |
| 2006 |  |  |  |  |  |  | 31 |  |
| 2007 |  |  |  |  |  |  | 39 |  |
| 2008 | 8 | 6 | 12 | 7 | 7 | 11 | 47 | 98 |
| 2009 | 22 | 11 | 25 | 20 | 6 | 13 | 48 | 145 |
| 2010 | 19 | 14 | 24 | 9 | 6 | 10 | 40 | 122 |
| 2011 | 20 | 11 | 21 | 12 | 7 | 14 | 25 | 110 |
| 2012 | 20 | 16 | 28 | 13 | 7 | 15 | 26 | 125 |
| 2013 | 25 | 12 | 22 | 14 | 5 | 14 | 28 | 120 |
| 2014 | 22 | 14 | 12 | 9 | 8 | 16 | 32 | 113 |
| 2015 | 26 | 11 | 24 | 12 | 8 | 14 | 36 | 131 |

Table 16.3.1.2 Cod abundance indices ('000) from the Greenland Shrimp and Fish survey by year and strata divisions in ICES 14.b and NAFO 1F. Q1 being the northern strata in East Greenland.

|  |  | ICES 14.B |  |  | NAFO |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Q1 | Q2 | Q3 | Q4 | Q5 | Q6 | 1F | Total | CV |
| 1992 |  |  |  |  |  |  | 8 |  |  |
| 1993 |  |  |  |  |  |  | 18 |  |  |
| 1994 |  |  |  |  |  |  | 0 |  |  |
| 1995 |  |  |  |  |  |  | 39 |  |  |
| 1996 |  |  |  |  |  |  | 107 |  |  |
| 1997 |  |  |  |  |  |  | 0 |  |  |
| 1998 |  |  |  |  |  |  | 3 |  |  |
| 1999 |  |  |  |  |  |  | 0 |  |  |
| 2000 |  |  |  |  |  |  | 189 |  |  |
| 2001 |  |  |  |  |  |  | 313 |  |  |
| 2002 |  |  |  |  |  |  | 457 |  |  |
| 2003 |  |  |  |  |  |  | 211 |  |  |
| 2004 |  |  |  |  |  |  | 1610 |  |  |
| New survey Gear introduced |  |  |  |  |  |  |  |  |  |
| 2005 |  |  |  |  |  |  | 86410 |  |  |
| 2006 |  |  |  |  |  |  | 39475 |  |  |
| 2007 |  |  |  |  |  |  | 32575 |  |  |
| 2008 | 5456 | 1361 | 13043 | 1975 | 1635 | 7958 | 22887 | 54314 | 22 |
| 2009 | 14304 | 2191 | 28539 | 4374 | 548 | 4753 | 1776 | 56486 | 15 |
| 2010 | 5844 | 732 | 30042 | 3975 | 115 | 4633 | 6557 | 51897 | 45 |
| 2011 | 7843 | 1357 | 5178 | 7733 | 1470 | 19072 | 6330 | 48983 | 22 |
| 2012 | 5475 | 2164 | 3658 | 2453 | 352 | 8635 | 21238 | 43975 | 20 |
| 2013 | 11102 | 1420 | 5667 | 17360 | 537 | 27145 | 49874 | 113104 | 32 |
| 2014 | 4168 | 3445 | 2622 | 19267 | 493 | 5412 | 22702 | 58106 | 36 |
| 2015 | 6396 | 4074 | 6941 | 3093 | 231 | 8322 | 34032 | 63090 | 28 |

Table 16.3.1.3. Cod biomass indices (tons) from the Greenland Shrimp and Fish survey by year and strata divisions in ICES 14.b (Q1-Q6) and NAFO 1F. Smoothed index is a random effects survey smoother applied to the total index.


Table 16.3.1.4: Abundance indices ('000) by age from the Greenland Shrimp and Fish survey by year in ICES 14.b + NAFO 1F.

| East Greenland |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year/age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 2008 | 4355 | 372 | 1113 | 7968 | 6582 | 23794 | 5412 | 2235 | 736 | 1006 | 739 |
| 2009 | 14970 | 7642 | 8019 | 4504 | 5378 | 5664 | 6610 | 2537 | 225 | 554 | 385 |
| 2010 | 150 | 2436 | 3959 | 5759 | 3253 | 12785 | 7969 | 11264 | 2958 | 450 | 914 |
| $2011$ | 315 | 162 | 5682 | 8288 | 16346 | 5409 | 4707 | 2226 | 3382 | 1834 | 634 |
| 2012 | 0 | 258 | 1208 | 12748 | 7154 | 12041 | 4155 | 2428 | 1345 | 1849 | 790 |
| 2013 | 0 | 157 | 1432 | 1954 | 44843 | 25373 | 26654 | 5209 | 3440 | 1852 | 2190 |
| 2014 | 692 | 15 | 207 | 1849 | 1558 | 21863 | 8805 | 12411 | 2875 | 3790 | 4041 |
| 2015 | 0 | 86 | 38 | 1259 | 4916 | 11445 | 29010 | 7407 | 4793 | 1954 | 2181 |

Table 16.3.1.5 The abundance indices ('000) by year class/age from the Greenland Shrimp and Fish survey subareas in ICES 14.b and NAFO 1F, 2015.

| YEAR CLASS | $\mathbf{2 0 1 5}$ | $\mathbf{2 0 1 4}$ | $\mathbf{2 0 1 3}$ | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 1}$ | $\mathbf{2 0 1 0}$ | $\mathbf{2 0 0 9}$ | $\mathbf{2 0 0 8}$ | 2007 | $\mathbf{2 0 0 6}$ | $<2006$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| ICES Q1 | 0 | 0 | 0 | 173 | 250 | 187 | 869 | 1193 | 1442 | 1019 | 1262 |
| ICES Q2 | 0 | 6 | 0 | 1 | 58 | 537 | 1620 | 974 | 514 | 214 | 150 |
| ICES Q3 | 0 | 0 | 0 | 266 | 769 | 531 | 1158 | 1628 | 1671 | 428 | 491 |
| ICES Q4 | 0 | 42 | 0 | 59 | 118 | 208 | 858 | 795 | 511 | 225 | 277 |
| ICES Q5 | 0 | 7 | 0 | 54 | 0 | 0 | 70 | 14 | 80 | 7 | 0 |
| ICES Q6 | 0 | 31 | 0 | 40 | 457 | 819 | 4805 | 1718 | 391 | 61 | 0 |
| NAFO 1F | 0 | 0 | 38 | 667 | 3264 | 9162 | 19630 | 1085 | 185 | 0 | 0 |

Table 16.3.2.1 German survey. Numbers of valid hauls by stratum in South and East Greenland, stratum 9 furthest to the north.

|  | NAFO 1 F |  | ICES 14.8 |  |  |  |  | STR |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | STR | STR | STR | STR | STR | STR |  |  |
| Year | STR 4.1 | 4.2 | 5.1 | 5.2 | 7.1 | 7.2 | 8.2 | 9.2 | Sum |
| 1981 | 1 | 2 | 2 | 12 | 4 | 12 | 19 | 10 | 62 |
| 1982 | 13 | 2 | . | 12 | 1 | 9 | 15 | 15 | 67 |
| 1983 | 18 | 4 | 1 | 26 | 8 | 14 | 25 | 10 | 106 |
| 1984 | 20 | 4 | 4 | 5 | 1 | 5 | 7 | 2 | 48 |
| 1985 | 21 | 4 | 5 | 22 | 11 | 26 | 35 | 18 | 142 |
| 1986 | 20 | 3 | 2 | 27 | 11 | 14 | 31 | 34 | 142 |
| 1987 | 21 | 5 | 16 | 25 | 7 | 21 | 26 | 11 | 132 |
| 1988 | 18 | 2 | 20 | 19 | 10 | 13 | 36 | 9 | 127 |
| 1989 | 25 | 3 | 37 | . | 20 | . | 26 | 4 | 115 |
| 1990 | 21 | 6 | 15 | 24 | 4 | 6 | 15 | 12 | 103 |
| 1991 | 14 | 5 | 9 | 18 | 11 | 7 | 45 | 13 | 122 |
| 1992 | 7 | 5 | . | . | . | . | 4 | 2 | 18 |
| 1993 | 7 | . | 9 | 9 | 5 | 5 | 15 | 10 | 60 |
| 1994 | 7 | 5 | . | . | . | . | . | 6 | 18 |
| 1995 | 10 | 5 | 8 | 8 | 5 | 4 | 16 | 8 | 64 |
| 1996 | 10 | 5 | 7 | 9 | 5 | 3 | 13 | 6 | 58 |
| 1997 | 8 | 5 | 5 | 6 | 4 | 1 | 9 | 5 | 43 |
| 1998 | 10 | 5 | 5 | 9 | 6 | 2 | 12 | 6 | 55 |
| 1999 | 9 | 3 | 5 | 7 | 4 | 4 | 10 | 6 | 48 |
| 2000 | 9 | 5 | 6 | 7 | 8 | 4 | 12 | 9 | 60 |
| 2001 | 11 | 6 | 5 | 8 | 8 | 2 | 17 | 12 | 69 |
| 2002 | 8 | 4 | 6 | 7 | 5 | 2 | 10 | 7 | 49 |
| 2003 | 7 | 5 | 5 | 5 | 5 | 1 | 12 | 10 | 50 |
| 2004 | 9 | 5 | 7 | 7 | 8 | 3 | 13 | 11 | 63 |
| 2005 | 6 | 5 | 6 | 7 | 8 | 4 | 12 | 9 | 57 |
| 2006 | 8 | 5 | 3 | 1 | 5 | 4 | 11 | 7 | 44 |
| 2007 | 9 | 5 | 4 | 6 | 4 | 3 | 13 | 8 | 52 |
| 2008 | 7 | 6 | 6 | 8 | 4 | 3 | 10 | 8 | 52 |
| 2009 | 5 | 5 | 2 | 5 | 5 | 4 | 9 | 8 | 43 |
| 2010 | 10 | 6 | 1 | 3 | 8 | 3 | 14 | 8 | 53 |
| 2011 | 6 | 6 | 5 | 8 | 6 | 4 | 14 | 9 | 58 |
| 2012 | 10 | 6 | 6 | 7 | 8 | 3 | 12 | 9 | 61 |
| 2013 | 9 | 6 | 5 | 9 | 7 | 5 | 15 | 9 | 65 |
| 2014 | 10 | 6 | 5 | 7 | 10 | 6 | 20 | 11 | 75 |
| 2015 | 8 | 6 | 6 | 8 | 9 | 10 | 19 | 9 | 75 |

Table 16.3.2.2 German survey. Cod abundance indices ('000) from the German survey in South and East Greenland by year and stratum.

| NAFO 1 F |  |  |  |  | ICES $14 . \mathrm{B}$ |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | STR4_1 | STR4_2 | STR5_1 | STR5_2 | STR7_1 | STR7_2 | STR8_2 | STR9_2 | Sum | SD |
| 1982 | 8540 | 1245 | . | 366 | 297 | 1493 | 664 | 385 | 12990 | 4973 |
| 1983 | 5267 | 2870 | 209 | 715 | 149 | 564 | 529 | 726 | 11029 | 3796 |
| 1984 | 3296 | 42 | 1268 | 413 | 138 | 750 | 173 | 333 | 6413 | 3845 |
| 1985 | 3492 | 1164 | 920 | 166 | 560 | 1554 | 401 | 310 | 8567 | 1978 |
| 1986 | 8967 | 492 | 3509 | 359 | 776 | 2641 | 1207 | 337 | 18288 | 5097 |
| 1987 | 23219 | 306 | 5655 | 4145 | 399 | 6298 | 1293 | 234 | 41549 | 14816 |
| 1988 | 28259 | 17 | 2590 | 2073 | 302 | 1175 | 738 | 601 | 35755 | 16719 |
| 1989 | 31810 | 31442 | 9979 | . | 880 | . | 2128 | 639 | 76878 | 42682 |
| 1990 | 7052 | 6306 | 2808 | 1155 | 861 | 4295 | 2799 | 468 | 25744 | 7720 |
| 1991 | 1367 | 233 | 790 | 937 | 122 | 368 | 652 | 510 | 4979 | 1548 |
| 1992 | 113 | 134 | . | . | . | . | 228 | 367 | 842 | 192 |
| 1993 | 0 | . | 613 | 62 | 127 | 317 | 114 | 148 | 1381 | 521 |
| 1994 | 44 | 12 | . | . | . | . | . | 234 | 290 | 135 |
| 1995 | 27 | 8 | 89 | 25 | 450 | 3082 | 77 | 91 | 3849 | 1314 |
| 1996 | 156 | 0 | 109 | 0 | 37 | 279 | 29 | 160 | 770 | 173 |
| 1997 | 49 | 0 | 25 | 17 | 200 | 54 | 145 | 1107 | 1597 | 479 |
| 1998 | 40 | 8 | 97 | 0 | 57 | 57 | 24 | 266 | 549 | 142 |
| 1999 | 155 | 0 | 198 | 8 | 165 | 1267 | 116 | 105 | 2014 | 582 |
| 2000 | 76 | 13 | 348 | 15 | 431 | 180 | 25 | 143 | 1231 | 251 |
| 2001 | 343 | 3 | 319 | 27 | 309 | 299 | 204 | 1071 | 2575 | 544 |
| 2002 | 1739 | 0 | 116 | 273 | 769 | 459 | 186 | 875 | 4417 | 1352 |
| 2003 | 840 | 8 | 199 | 183 | 1250 | 1399 | 1100 | 1438 | 6417 | 1004 |
| 2004 | 10902 | 107 | 1684 | 133 | 285 | 1817 | 1401 | 1073 | 17402 | 8499 |
| 2005 | 24438 | 1399 | 16577 | 3078 | 718 | 7157 | 1580 | 2070 | 57017 | 11411 |
| 2006 | 28894 | 486 | 14733 | 3686 | 6044 | 7378 | 2779 | 2700 | 66700 | 15653 |
| 2007 | 67049 | 772 | 2283 | 3256 | 758 | 5363 | 2080 | 2093 | 83654 | 56843 |
| 2008 | 18730 | 292 | 2036 | 4898 | 2203 | 9460 | 1285 | 2678 | 41582 | 10268 |
| 2009 | 1286 | 283 | 1017 | 567 | 3129 | 8755 | 1566 | 3275 | 19878 | 3581 |
| 2010 | 2372 | 141 | 532 | 1703 | 1101 | 8875 | 933 | 1748 | 17405 | 2958 |
| 2011 | 7547 | 162 | 3027 | 1326 | 868 | 1971 | 1243 | 2816 | 18960 | 3196 |
| 2012 | 23964 | 132 | 5689 | 167 | 901 | 2117 | 1114 | 3982 | 38066 | 22168 |
| 2013 | 41722 | 1947 | 2193 | 818 | 874 | 3121 | 1157 | 1342 | 53174 | 43105 |
| 2014 | 73612 | 111 | 8612 | 4013 | 228 | 1089 | 1436 | 5461 | 94562 | 77704 |
| 2015 | 3187 | 361 | 1186 | 267 | 113 | 834 | 2265 | 3395 | 11608 | 3752 |

Table 16.3.2.3 German survey. Cod biomass indices (tons) from the German survey in South and East Greenland by year and stratum.

| NAFO 1 F |  |  |  |  | ICES $14 . \mathrm{B}$ |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | STR4_1 | STR4_2 | STR5_1 | STR5_2 | STR7_1 | STR7_2 | STR8_2 | STR9_2 | Sum | SD |
| 1982 | 14607 | 3690 | . | 1201 | 1036 | 3342 | 2576 | 1900 | 28352 | 8415 |
| 1983 | 9797 | 6219 | 653 | 2209 | 402 | 2294 | 2605 | 4442 | 28621 | 8201 |
| 1984 | 5326 | 82 | 3115 | 1444 | 346 | 1782 | 540 | 2553 | 15188 | 6650 |
| 1985 | 2942 | 1976 | 1812 | 803 | 1393 | 3875 | 1187 | 1605 | 15593 | 3099 |
| 1986 | 8005 | 943 | 1044 | 873 | 2537 | 3921 | 2301 | 709 | 20333 | 6054 |
| 1987 | 17186 | 276 | 2889 | 3735 | 504 | 10243 | 4558 | 1414 | 40805 | 16521 |
| 1988 | 26349 | 17 | 2812 | 4605 | 964 | 2297 | 3475 | 2012 | 42531 | 18651 |
| 1989 | 36912 | 35281 | 23605 | . | 2518 | . | 6889 | 2174 | 107379 | 61579 |
| 1990 | 9212 | 5897 | 5361 | 3215 | 2517 | 10386 | 6551 | 1620 | 44759 | 10905 |
| 1991 | 2088 | 200 | 1465 | 2759 | 196 | 1008 | 2610 | 2100 | 12426 | 4657 |
| 1992 | 79 | 50 | . | . | . | . | 171 | 734 | 1034 | 286 |
| 1993 | 0 | . | 431 | 73 | 247 | 532 | 254 | 547 | 2084 | 588 |
| 1994 | 2 | 7 | . | . | - | . | . | 779 | 788 | 514 |
| 1995 | 6 | 4 | 32 | 62 | 166 | 11744 | 250 | 123 | 12387 | 5550 |
| 1996 | 101 | 0 | 63 | 0 | 109 | 708 | 99 | 511 | 1591 | 333 |
| 1997 | 53 | 0 | 18 | 20 | 358 | 70 | 337 | 4017 | 4873 | 1800 |
| 1998 | 12 | 11 | 29 | 0 | 87 | 122 | 123 | 986 | 1370 | 554 |
| 1999 | 39 | 0 | 24 | 1 | 162 | 2229 | 492 | 201 | 3148 | 1184 |
| 2000 | 13 | 9 | 132 | 17 | 206 | 616 | 75 | 540 | 1608 | 366 |
| 2001 | 88 | 5 | 130 | 19 | 345 | 382 | 387 | 3005 | 4361 | 1593 |
| 2002 | 976 | 0 | 38 | 224 | 1547 | 531 | 541 | 2214 | 6071 | 1306 |
| 2003 | 361 | 17 | 121 | 266 | 3787 | 2440 | 1716 | 4169 | 12877 | 2817 |
| 2004 | 1945 | 177 | 359 | 55 | 957 | 2319 | 3264 | 3240 | 12316 | 3070 |
| 2005 | 9055 | 1870 | 8135 | 2537 | 3155 | 17882 | 3590 | 6806 | 53030 | 7772 |
| 2006 | 31616 | 681 | 8616 | 4130 | 3557 | 10291 | 6084 | 11567 | 76542 | 24680 |
| 2007 | 74671 | 1045 | 3749 | 5042 | 1363 | 14456 | 5374 | 8540 | 114240 | 58452 |
| 2008 | 18543 | 344 | 3630 | 9790 | 5075 | 26506 | 3772 | 11908 | 79568 | 12433 |
| 2009 | 583 | 277 | 1361 | 1726 | 10145 | 28613 | 6351 | 15520 | 64576 | 13358 |
| 2010 | 3629 | 273 | 741 | 5085 | 5244 | 31745 | 4282 | 10932 | 61931 | 11626 |
| 2011 | 12398 | 385 | 5839 | 4364 | 1658 | 8051 | 5735 | 17487 | 55917 | 10240 |
| 2012 | 33871 | 370 | 15679 | 579 | 2596 | 6245 | 5445 | 26885 | 91670 | 30054 |
| 2013 | 74193 | 6525 | 6672 | 2737 | 2577 | 9752 | 4853 | 7575 | 114884 | 75148 |
| 2014 | 132706 | 428 | 31885 | 15935 | 1060 | 4322 | 6480 | 29358 | 222174 | 132209 |
| 2015 | 11848 | 1534 | 3938 | 1804 | 522 | 3645 | 9891 | 19119 | 52301 | 16354 |

Table 16.3.2.4 German survey, South and East Greenland (NAFO 1F and ICES 14.). Age disaggregate abundance indices ('1000).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | $11+$ | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 |  | 23 | 214 | 2500 | 1760 | 4451 | 1952 | 793 | 223 | 927 | 57 | 74 | 12974 |
| 1983 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1984 | 23 | 8 | 54 | 1134 | 507 | 2434 | 582 | 1242 | 229 | 125 | 17 | 49 | 6404 |
| 1985 | 279 | 2521 | 242 | 160 | 1658 | 947 | 1439 | 344 | 831 | 96 | 27 | 27 | 8571 |
| 1986 |  | 3367 | 9255 | 1128 | 273 | 1631 | 603 | 1300 | 165 | 473 | 31 | 58 | 18284 |
| 1987 |  | 4 | 10193 | 24656 | 2689 | 720 | 1368 | 296 | 966 | 80 | 487 | 49 | 41508 |
| 1988 | 6 | 18 | 335 | 9769 | 23391 | 876 | 200 | 559 | 83 | 337 | 31 | 146 | 35751 |
| 1989 | 12 | 2 | 111 | 732 | 23945 | 49864 | 1007 | 44 | 756 | 70 | 282 | 76 | 76901 |
| 1990 | 58 | 36 | 58 | 715 | 706 | 11679 | 12101 | 139 | 15 | 74 |  | 148 | 25729 |
| 1991 |  | 73 | 150 | 171 | 539 | 102 | 2128 | 1762 | 31 | 11 | 3 | 9 | 4979 |
| 1992 | 214 | 10 | 196 | 103 | 61 | 53 | 67 | 67 | 51 |  |  | 21 | 822 |
| 1993 |  | 4 | 15 | 869 | 152 | 95 | 97 | 31 | 83 | 34 |  | 2 | 1382 |
| 1994 |  | 71 | 5 | 16 | 84 | 39 | 22 | 38 |  | 8 |  | 0 | 283 |
| 1995 |  | 1 | 621 | 347 | 260 | 1399 | 372 | 120 | 403 | 32 | 192 | 102 | 3849 |
| 1996 |  | 0 | 0 | 353 | 130 | 131 | 110 | 23 | 25 |  |  | 0 | 772 |
| 1997 |  | 0 | 12 | 17 | 687 | 557 | 191 | 78 | 48 |  |  | 5 | 1595 |
| 1998 | 51 | 73 | 39 | 4 | 11 | 173 | 138 | 48 | 10 |  |  | 0 | 547 |
| 1999 | 105 | 426 | 389 | 346 | 118 | 257 | 174 | 156 |  | 29 | 16 | 0 | 2016 |
| 2000 |  | 202 | 243 | 323 | 208 | 40 | 72 | 20 | 46 | 61 | 15 | 0 | 1230 |
| 2001 |  | 166 | 568 | 493 | 631 | 362 | 190 | 60 | 50 | 18 | 10 | 2 | 2550 |
| 2002 | 40 | 1 | 395 | 2119 | 601 | 477 | 454 | 217 | 61 | 21 | 11 | 7 | 4404 |
| 2003 | 579 | 629 | 53 | 553 | 1761 | 1026 | 1015 | 541 | 220 | 37 | . | 4 | 6418 |
| 2004 | 386 | 10687 | 1770 | 448 | 617 | 1667 | 921 | 620 | 228 | 39 | 10 | 8 | 17401 |
| 2005 | 80 | 1603 | 39549 | 8091 | 1250 | 2819 | 2549 | 727 | 189 | 40 |  | 0 | 56897 |
| 2006 | 80 | 439 | 3375 | 48140 | 9269 | 1328 | 2404 | 1309 | 193 | 30 | 9 | 0 | 66576 |
| 2007 | 128 | 154 | 2007 | 5149 | 65974 | 8166 | 713 | 658 | 634 | 70 |  | 0 | 83653 |
| 2008 | 14 | 265 | 513 | 8213 | 4401 | 22939 | 4201 | 516 | 220 | 199 | 44 | 29 | 41554 |
| 2009 | 98 | 322 | 1057 | 391 | 1620 | 2863 | 11241 | 1964 | 111 | 134 | 64 | 17 | 19882 |
| 2010 | 22 | 700 | 1425 | 1388 | 845 | 2887 | 2518 | 5707 | 1362 | 236 | 163 | 139 | 17392 |
| 2011 |  | 120 | 1246 | 3475 | 4874 | 2402 | 2949 | 1179 | 2324 | 310 | 23 | 49 | 18951 |
| 2012 | 6 | 50 | 1624 | 10093 | 10233 | 9846 | 2827 | 1778 | 1166 | 379 | 35 | 5 | 38042 |
| 2013 |  | 17 | 35 | 4312 | 27014 | 11146 | 7455 | 1314 | 517 | 291 | 126 | 68 | 52295 |
| 2014 |  | 7 | 55 | 602 | 20847 | 58174 | 9275 | 3284 | 1316 | 494 | 441 | 52 | 94547 |
| 2015 | 105 | 37 | 68 | 341 | 752 | 3688 | 3598 | 1881 | 644 | 187 | 106 | 160 | 11567 |

Table 16.5.1. Number of tagged cod in the period of 2003 to 2015 in different regions Fjord (West) = NAFO divisions 1B-1F. Bank (West) = NAFO division 1C+1D+1E. East Greenland = NAFO division 1F + ICES division 14.b.

| YEAR | FJORD | BANK (WEST) | EAST GREENLAND |
| :---: | :---: | :---: | :---: |
| 2003 | 599 |  |  |
| 2004 | 658 |  | 1387 |
| 2005 | 565 | 721 | 1296 |
| 2006 | 41 |  | 525 |
| 2007 | 1140 |  | 403 |
| 2008 | 231 |  | 2359 |
| 2009 | 633 | 1563 | 1203 |
| 2010 | 88 | 2321 | 1218 |
| 2011 | 28 |  |  |
| 2012 | 86 | 57 |  |
| 2014 |  |  |  |

Table 16.5.2: Number of recaptured cod in the period of 2003 to 2015 in different regions Fjord = NAFO divisions 1B-1F. Bank (West) = NAFO division 1D+1E. East Greenland = NAFO division 1F + ICES division 14.b.

|  | FJord | Bank (West) | East Greenland |
| :--- | :---: | :---: | :---: |
| Fjord (West) | 438 | 14 | 2 |
| Bank (West) | 1 | 34 | 2 |
| East Greenland |  | 12 | 99 |
| Iceland | 3 | 21 | 125 |

### 16.13Figures



Figure. 16.1. Sampling location of spawning cod in Greenland and Iceland in the genetic project. The colours of the dots represent the blends of sample mean of the different spawning population: West offshore, Nuuk (inshore), East (Greenland and offshore Iceland) and Iceland inshore as signal intensities of green and red respectively. After Therkildsen et al. 2013.


Figure 16.2.1. Annual total catch in South and East Greenland (NAFO subarea 1F and ICES Subarea 14.b). From 2001 divided into gear. TAC until 2013 is for all the offshore area including West Greenland (NAFO subarea 1A-1E).


Figure 16.2.2.1: Annual distribution of total catches of Atlantic cod in West and East Greenland. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 16.2.2.1: Continued. Annual distribution of total catches of Atlantic cod in West and East Greenland. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 16.2.2.2: Distribution of Longline and Trawl catches of Atlantic cod in West and East Greenland 2015. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 16.2.3.1: Combined length and age distributions of commercial cod catches in the South and East Greenland offshore fishery in 2015.


Figure 16.2.3.2: Length and age distributions of commercial cod catches in the four management areas of SouthWest (NAFO 1F) and East Greenland (Q1Q2 furthest north) in 2015.


Figure 16.2.3.3: Catch-at-age in the East Greenland (ICES 14b + NAFO 1F) commercial fishery. Size of circles represents size of catch numbers.


Figure 16.2.4.1: Ln cpue (ton/h) for Atlantic Cod caught in the fishery in East (ICES 14b) and SouthWest (NAFO 1F) Greenland. Based on model: In cpue = year + management area (Q1Q2, Q3Q4, Q5Q6 and 1F) + ship. Dashed lines are 2*SE.


Figure16.3.1.1. Greenland shrimp and fish survey 2008-2015. Abundance per $\mathrm{Km}^{2}$.


Figure 16.3.1.2. Greenland shrimp and fish survey 2008-2015. Catch weight kg per $\mathrm{Km}^{2}$


Figure 16.3.1.3: Abundance index pr. age in ICES 14b and NAFO 1F combined. Size of circles represents size of index.


Figure 16.3.1.4: Abundance index by length (cm) and area in 2015. Areas from north (top) to south (bottom) is: Q1, Q2, Q3, Q4, Q5, Q6 (ICES 14b) and NAFO 1F.


Figure 16.3.1.5. Abundance ( $\mathrm{no} / \mathrm{km}^{2}$ ) pr. station of ages 1-10 in the years 2009-2015.


Figure 16.3.1.6: Total abundance indices by length in East Greenland (ICES 14b + NAFO 1F) shrimp and fish survey.


Figure 16.3.1.7: Estimated SSB (tons) by year from the East Greenland (ICES 14b + NAFO 1F) Shrimp and Fish survey.


Figure 16.3.1.8: Estimated SSB (tons) by survey areas from the East Greenland (ICES 14b + NAFO 1F) Shrimp and Fish survey, 2015. NAFO Div 1F (SouthWest Greenland) to the left, " $Q$ " areas (East Greenland) to the right. Cape Farewell is between 1F and Q6.


Figure 16.3.1.9: Biomass index for NAFO 1F and ICES Subarea 14b. Red squares are the estimated mean value from the survey and the vertical connected lines are upper and lower $95 \%$ confidence intervals. The smoothed estimates are displayed as the blue line and the $95 \%$ confidence intervals of the smoothed values are shown as dashed lines.

Abundance distr. Atlantic cod, 2015


Figure 16.3.2.1 German survey, 2015. Abundance (num per km2) pr haul.

Biomass distr. Atlantic cod, 2015


Figure 16.3.2.2 German survey, 2015. Biomass (kg per km2) pr haul.


Figure 16.3.2.3 German survey, Cod off Greenland. Abundance per age group and stratum 2015. Strata 1-4 is West Greenland from north to south; strata 5-9 is East Greenland from south to north.


Figure 16.3.2.4 German survey, Cod off Greenland. Abundance indices for South and East Greenland.


Figure 16.3.2.5 German survey, Cod off Greenland. Biomass indices for South and East Greenland.


Figure 16.10.1: Catch, catch advice with different cap's on the 3.2 DLS approach and German survey index (without scale). The two figures show the same, but with two different scales on the Y-axis.


Figure 16.10.2: Catch (black), catch advice with setoff in year 1994 (blue), catch advice with setoff in year 1997 (green). Both $20 \%$ cap and 0.8 buffer were used for the two catch-advice calculations.


Figure 16.10.3: Log catch ratios from the commercial catches fitted with a Loess smoother. Labels are age specific $\log$ catch ratios in a given year.

## 17 Greenland Halibut in Subareas 5, 6, 12, and 14

Greenland halibut in ICES Subareas 5, 6, 12 and 14 are assessed as one stock unit although precise stock associations are not known.

### 17.1 Executive summary

Catches of Greenland halibut in Subareas 5,6,12 and 14 have ranged between 20 and 30 kt in the last two decades and amounted 25 kt in 2015.The biomass indices used as input to the assessment (combined survey index at Greenland and Iceland) and logbook information from Iceland trawler fishery all show a slight increase in 2015.,

A logistic production model in a Bayesian framework has been used to assess stock status and for making predictions. The model includes an extended catch series going back to the assumed virgin status of the stock at the beginning of the fishery in 1961. Estimated stock biomass showed an overall decline along with the high catches in the late 1980s and early 1990s. Since 2004/2005 the stock has increased slowly and is now at $71 \% B_{\text {MSY }}$.Fishing mortality have since 2013 been around $\mathrm{F}_{\text {MSY }}$ and is in 2015 10\% above $\mathrm{F}_{\text {msy }}$. The remaining available indices that are not currently used in the analytical assessment, e.g. logbook from East Greenland trawl fishery and from Faroese trawl fishery and a Faroese survey suggest high biomass in recent years, and therefore supports the recent trend in the assessment.

### 17.2 Catches, Fisheries, Fleet and Stock Perception

### 17.2.1 Catches

Total annual catches in Divisions 5a, 5b, and Subareas 6, 12 and 14 are presented for the years 1981-2015 in Tables 17.2.1-17.2.6 and since 1961 in Figure 17.2.1. Catches increased in 2015 by $22 \%$ to 256.67 t. Landings in Icelandic waters (usually allocated to Division 5a) have historically predominated the total landings in areas $5+14$, but since the mid 1990s also fisheries in Subarea 14 and Division 5b have developed. Landings have since 1997 been between 20-31 kt.

### 17.2.2 Fisheries and fleets

In 2015 quotas in Greenland EEZ and Iceland EEZ were fully utilized as in the preceding fishing years. In the Faroe EEZ the fishery is regulated by a fixed numbers of licenses and technical measures like bycatch regulations for the trawlers and depth and gear restrictions for the gillnetters.

Most of the fishery for Greenland halibut in Divisions 5a, 5b and $14 b$ is a directed trawl fishery, and also an insignificant gillnet ( $\sim 5 \%$ in 2015) and longline fishery ( $\sim 1 \%$ in 2015) takes place. Only minor catches in 5 a and 14 b are taken as bycatches in a redfish fishery (see section 23 on Greenland slope redfish). No or insignificant discarding has been observed in this fishery.

Spatial distribution of 2015 fishery and historic effort and catch in the trawl fishery in Subareas 5, 6, 12 and 14 is provided in Figures 17.2.2-5. Fishery in the entire area did in the past occur in a more or less continuous belt on the continental slope from the slope of the Faroe plateau to southeast of Iceland extending north and west of Iceland and further south to southeast Greenland. Fishing depth ranges from $350-500 \mathrm{~m}$ southeast, east and north of Iceland to about 1500 m at East Greenland. In 2015 the distribution of the fishery covered all areas but was discontinuous in its distribution.

In 2001-2008 a directed and a bycatch fishery by Spain, France, Lithuania, UK and Norway developed in the Hatton-bank area of Division 6b. However, most of these fisheries ceased after 2008 and is currently insignificant. Landings in Subareas 12 and 6 in Tables 17.2.5-17.2.6 derive from the Hatton-bank area.

### 17.2.3 Bycatch and discard

The Greenland halibut trawl fishery is commonly a clean fishery with respect to bycatches. Eventual bycatches are mainly redfish and cod. Southeast of Iceland the cod fishery and a minor Greenland halibut fishery are coinciding spatially. In East Greenland where fishery is on the steep slope, fishing grounds for cod and redfish are close to the Greenland halibut fishing grounds, but nevertheless the catches from single hauls are clean.

The mandatory use of sorting grids in the shrimp fishery in Icelandic and Greenland waters since 2002 are observed to have reduced bycatches considerably. Based on sampling in 2006-2007, scientific staff observed bycatches of Greenland halibut to be less than $1 \%$ compared to about $50 \%$ by weight observed before the implementation of sorting grids (Sünksen 2007). No information has since been available but the fishery in 14 b generally report discard rates less than $1 \%$ by weight in logbooks.

### 17.3 Trends in Effort and cpue

### 17.3.1 Division 5a

Indices of cpue for the Icelandic trawl fleet directed at Greenland halibut for the period 1985-2015 is provided in Table 17.3.1 and Figures 17.3.1-3. At latest benchmark (WKBUT 2013) the cpue series from this fishery was questioned due to a marked distributional change in fishing season and area, and also because the regulations might have caused a changed behaviour in the fishing fleets. The important fishing grounds west of Iceland, where approximately $70-80 \%$ of the landings historically came from, are the areas where the season shift mainly has affected the cpue. A simple standardization procedure was not considered sufficient to account for these changes (Figure 17.3.2.). A rough estimate on stock biomass distribution in Iceland is an equal quantity in each of the four areas and the overall cpue index for the Icelandic fishery was therefore compiled as the average of the standardized indices from the four areas (Fig 17.3.12.).

Catch rates of Icelandic bottom trawlers decreased for all fishing grounds during 19901996 (Figure 17.3.1) but have since peaked in 2001 and have in recent years been stable and slowly increasing. The overall tendency is the same for all fishing grounds in 5 a (Figure 15.3.2) although the less important fishing grounds in north, east and southeast more variable in trend. Both observed and derived effort are about historic average in 2014 (Figure 17.3.3).

### 17.3.2 Division 5b

Information from logbooks from the Faroese otterboard trawl fleet ( $>1000 \mathrm{hp}$ ) was available for the years 1991-2015 (Table 17.3.1, Figure 17.3.4.). The bulk of the fishery has historically been on the southeast slope of the Faroe Plateau. cpue decreased drastically in the early period by more than $50 \%$ coinciding with a significant increase in effort. In 2011 cpue increased sharply by more than $60 \%$ until 2013 but has since decreased to $50 \%$ of that level.

### 17.3.3 Division 14b

cpue and effort from logbooks in 14 are provided in Table 17.3.1 and Figure 17.3.5-6. Following a period with relatively low cpues in 1999-2004, catch rates have been variable but increasing and reached in 2015 a record high. It should be noted that cpue series from Divisions 5a, 5b and 14b have different trends over the time indicating that the populations/areas do have different dynamics.

### 17.3.4 Divisions 6 and 12b

Since 2001 a fishery developed in Divisions 6b and 12b in the Hatton-bank area, but in both divisions the recent catches are relatively small. Limited fleet information is available from this area (ICES WGDEEP).

### 17.4 Catch composition

Length compositions of catches from the commercial trawl fishery in Div. 5a are rather stable from year to year. In Figure 17.4.1 length distributions are shown since 1996 from the western area of Iceland, comprising the most important fishing grounds. Distributions are rather stable over the entire period Catch composition from all areas (5a,b and 14) by gear is provided for 2015 in Figure 17.4.2.

### 17.5 Survey information

The total surveyed area in 2015 for Greenland halibut in Divisions 5a and 14b is provided in Figure 17.5.1. The areas where commercial fishing takes place (Figure 17.2.2.) are covered by the annual surveys. The two surveys in 5 a and 14 b are combined to one index and used as input in the assessment model.

### 17.5.1 Division 5a

Since 2006 the total biomass of Greenland halibut has increased significantly in Icelandic waters (Figures 17.5.3). (Figures 17.5.3.-17.5.4.). Abundance of smaller fish (less than 40 cm ) has been record low in recent two years indicating poor recruitment.

### 17.5.2 Division 5b

The catch rates from the available time-series of the Faroese survey have declined from a record high level in 2012 but is still high in 2015.(Figure 17.5.5).

### 17.5.3 Division 14b

Total biomass in the Greenlandic survey in 2015 was estimated 11285 tons (S.E. 3752) which is a $15 \%$ increase from the 2014 estimate. A GLM analysis performed on the survey catch rates, taking into account the scattered coverage of area and depth between years did however showed a slight decrease from 2014 (Figure 17.5.6.). The text table below provides information on the coverage and numbers of stations in 2015 along with the Iceland survey in Division 5a.

| SURVEY <br> /DIVISION | No. HAULS IN 2015 <br> (PLANNED HAULS) | DEPTH RANGE (M) | CoVERAGE (KM2) |
| :--- | :---: | :---: | :---: |
| 5 a | $203(219)$ | $32-1309 ?$ | -130000 |
| 14 b | $84(70)$ | $400-1500$ | 29000 |

The stock annex provides more extensive descriptions of the surveys.

### 17.6 Stock Assessment

### 17.6.1 .Stock production model

The assessment uses a stochastic version of the logistic production model and Bayesian inference according to the Stock Annex in which a more detailed formulation of the model and its performance is found.

### 17.6.1.1 Input data

The model synthesize information from input priors and two independent series of Greenland halibut biomass indices and one series of catches by the fishery (Table 17.6.1). The two series of biomass indices were: a revised and standardized series of annual commercial-vessel catch rates for 1985-2015, сриеt,; and a combined trawl-survey biomass index for 1996-2015, Isurt.

Total reported catch or WGs best estimates in ICES Subareas 5, 6, 12 and 14 1961-2015 was used as yield data (Table 17.6.1, Figure. 17.2.1). Since the fishery has no major discarding problems or misreporting, the reported catches were entered into the model as error-free.

### 17.6.1.2 Model performance

The model parameters were estimated (posterior) based on the prior assumptions (Table 17.6.2-3 and Figure 17.6.1). The data could not be expected to carry much information on the parameter $P_{1960}$ - the stock size 25 years prior to when the series of stock biomass series start - and the posterior resembled the prior (Figure 17.6.1). The prior for K was somewhat updated to slightly higher values. However, the posterior still had a wide distribution with an inter-quartile range of 701-1072 ktons (Table 17.6.3).

The model was able to produce a reasonable simulation of the observed data (Figure 17.6.2). The probabilities of getting more extreme observations than the realized ones given in the dataseries on stock size were in the range of 0.05 to 0.95 i.e. the observations did not lay in the extreme tails of their posterior distributions (Table 17.6.4). Exceptions are observed for the survey in 1997 ( $p=0.96$ ) and in 2006 ( $p=0.04$ ). The cpue series was generally better estimated than the survey series (Figure. 17.6.2).

The retrospective runs suggest high consistency (Figure. 17.6.3).

### 17.6.1.3. Assessment results

The time-series of estimated median biomass-ratios starts in 1960 as a virgin stock at K (Figure. 17.6.4-5). The fishery starts in 1961. Under continuously increasing fishing mortality the stock declined sharply in the mid-1990s to levels below the optimum, $B_{M S Y}$. Some rebuilding towards $B_{M S Y}$ was then seen in the late 1990s. Since then the stock started to increase from its lowest level in 2004-5 of approx. $45 \%$ of BMSY. In 2014 biomass was at $71 \%$ of $B_{M S Y}$. The risk of the biomass being below $B_{M S Y}$ in 2014 is $100 \%$ and
 exceeded $F_{M S Y}$ since the 1990s and estimated at 1.3 $F_{M S Y}$ in 2014. (Figure. 17.6.4 and 17.6.5). This parameter can only be estimated with relatively large uncertainty and the posteriors therefore also include values below $F_{M S \gamma}$. However, the probability that the $F$ has exceeded $F_{M S Y}$ is high for most of the series.

The posterior for MSY was positively skewed with upper and lower quartiles at 26 ktons and 39 ktons (Table 17.6.3). As mentioned above MSY was relatively insensitive to changes in prior distributions.

Within a one-year perspective the sensitivity of the stock biomass to alternative catch options seems rather low. This is due to the inertia of the model used (see annex) and the low growth rate of the population. Risk associated with five optional catch levels for 2016 are given in Table 17.6.5.
The risk trajectory associated with ten-year projections of stock development assuming a maintained annual catch in the entire period ranging from 0 to 30 ktons were investigated (Figure. 17.6.6.-7). The calculated risk is a result of the projected development of the stock and the increase in uncertainty as projections are carried forward. It must be noted that a catch scenario of a maintained constant catch over a decade without considering arrival of new biological information and advice is highly unrealistic.

Catches around 20 ktons are likely to lead to an increase in stock size and annual catches of 15 kt or less will result in a $50 \%$ probability of reaching BMSY within 10 years (Figure 17.6.6).

Scenarios of fixed levels of fishing mortality ratios within the range of 0.3 to 1.7 were conducted and are shown in Figure 17.6.8. Present biomass is above the MSY Btrigger ( $50 \%$ of $\mathrm{BMSY}_{\text {M }}$ ) and a fishery at $\mathrm{F}_{\text {MSY }}$ is then advised according the ICES MSY approach. Fishing at Fmsy will result in catches of 22 kt in 2016 (Figure 15.6 .8 panel D) and a stock size of $71 \%$ of BMsץ in 2016 (Table 17.6.5).

### 17.6.2 Short-term forecast and management options

Biomass scenarios at various catch options are provided in Table 17.6.5 and Figures 17.6.6-7. Catches below 30 kt is estimated to lead to an increase in biomass, while catches of 30 kt will remain biomass at current level over the next decade. Catches of 24 kt in 2017 will correspond to fishing at $\mathrm{F}_{\mathrm{msy}}$. This will result in a slight increase in biomass and risk of exceeding Flim will remain unchanged from 2016 (Table 17.6.5). At catches of 24 kt the biomass is not expected to reach BMSY within the next decade although biomass will increase over the period.

### 17.6.3 Reference points

Reference points were defined at the benchmark in 2013 (WKBUT) and further at WKMSYREF4 in 2015: Blim as $30 \% \mathrm{~B}_{\text {MSY }}$ corresponding to production reduced to $50 \%$ of its maximum. This is equivalent to the SSB-level (spawning-stock biomass) at $50 \% \mathrm{R}_{\max }$ (maximum recruitment). Greenland halibut is believed to be a slow growing species i.e. with relative low $r$ (intrinsic rate of increase). This means that even without fishery it would take some 10 years to rebuild the stock from $30 \% \mathrm{BMSY}$ to BMSY (calculated by setting $r=0.21$, the $75^{\text {th }}$ percentile)-but likely longer.
MSY Btrigger, the biomass level that triggers a deviation from FmSY advice, was defined as $50 \% \mathrm{Bmsy}_{\text {. }}$ Flim was defined as $1.7 \mathrm{~F}_{\mathrm{mSY}}$.

### 17.7 Exploratory assessment: Gadget

An exploratory stock assessment for Greenland halibut in 5, 6, 12, and 14 using the Gadget modelling framework was presented at the NWWG-2016 meeting. Gadget can be viewed as a general framework for utilizing all available data and as such can detect inconsistencies in the data that are often ignored in other model types which are built
on top of highly aggregated data and/or make stronger assumptions on the stock dynamics. In general the exploratory Gadget model did seem to capture the main trends in the data.

### 17.7.1 Input data

The data used in the model were sex and length dis-aggregated indices from the Icelandic and Greenland surveys (combined in one index). Length distributions from the Icelandic and Greenland trawler fleet. Data on sex-ratio from the Icelandic trawler fleet and preliminary length-at-age data from 5b in 2015 was used to estimate growth inside the model. Catches from Iceland, Greenland and the Faroe Islands are included in the model.

### 17.7.2 Model settings

The model time was 1980-2015, with recruitment estimated annually. Two stocks are defined in the model, females and males that have different growth rates. Recruitment to the two stocks is equal that is $50: 50$. Initial population sizes for the stock components were estimated at varying initial depletion to account for the different fishing pressure on the sexes.

In the model three fleets are defined, Icelandic, Greenland and Faroe Islands trawlers. The commercial fleet operations in $5 \mathrm{a}, 5 \mathrm{~b}$ and 14 b were implemented separately but had the same selection pattern. In the model, natural mortality is set at 0.1 for all ages. The age range in the model is from age 3 to 25 with 25 being a plus-group.

### 17.7.3 Likelihood components

The likelihood components in the assessment are listed in table 17.7.1. In all the model has 14 likelihood components but two of those are mainly for constraining the minimization routines (understocking and bounds).

Text Table: Components in Gadget model.

| Component | TYPE | Notes |
| :--- | :--- | :--- |
| Comm.sex.ratio | stockdistribution | Sex ratio from Icelandic trawlers |
| ldist.f.survey | catchdistribution | Female length distribution from the combined <br> survey |
| ldist.m.survey | catchdistribution | Male length distribution from the combined survey |
| comm.ldist | catchdistribution | Length distributions from the Icelandic trawl fleet |
| si.F.20-50 | surveyindices | Female abundance survey index for length 20 to 50 <br> cm |
| surveyindices | Female abundance survey index for length 50 to 60 <br> and 60 to 70 cm |  |
| si.F.70-120 | surveyindices | Female abundance survey index for length larger <br> than 70 cm, split into 10 cm length bins |
| si.M.20-50 | surveyindices | Male abundance survey index for length 20 to 50 <br> cm |
| si.M.50-70 | surveyindices | Male abundance survey index for length 50 to 60 <br> and 60 to 70 cm |
| male abundance survey index for length 70 to 80 |  |  |
| and $80+$ |  |  |


| aldist.M.faroes | catchdistribution | Male age-length distribution from the Faroese <br> trawl survey in 2015 |
| :--- | :--- | :--- |
| understocking | understocking | To constrain minimization so that the stock will <br> always be larger than the catches |
| Bounds | bounds | To constrain minimization to respect the bounds of <br> the parameters. |

### 17.7.4 Fit to data

In general the model captures the changes in the length distributions from both from commercial catches and the survey. However the model has problems with estimating the proportions of males and females from commercial catches, where the proportion of large males in the catch is overestimated. The fit to the survey indices is mostly reasonable, the model following the main trends in them. Weights to individual likelihood components were assigned using an iterative reweighting heuristic described by Taylor et. al (2007).

### 17.7.5 Model estimates

Model estimates of spawning biomass, fishing mortality and recruitment are shown in figure 17.7.1. These biomass estimates follow similar patterns as expected, the spawn-ing-stock biomass has been decreasing since 1985, to an all-time low in 2013. The recruitment indices, however, appear more sporadic which is believed, based on experience with similar models, to related to a lack of reliable information on age for at least three years. As noted by Taylor et. al (2007) this could result in a level change in the size of the biomass, hence reference points and assumptions on stock productivity could therefore change. As illustrated at WKBUT-2013, the range of SSB and harvestable biomass estimates differed considerably based on different assumption on growth, whereas the advice, based on F0.1, varied between 18 to 30 kt for 2014 with a base case advice around 20 kt .

Figure 17.7.2 illustrates the equilibrium catch and spawning-stock biomass as a function of harvest rate estimated using stochastic forward simulations. In the simulations the $\mathrm{F}_{\text {msy }}$ was estimated at 0.27 , which was associated with the fishing mortality that has less $5 \%$ chance of going below $B \lim , B_{\text {lim set }}$ set the estimate of the lowest spawning-stock biomass. For reference the harvest rates associated with $\mathrm{F}_{0.1}, \mathrm{~F}_{\mathrm{pa}}, \mathrm{F}_{\mathrm{mSY}}$ and $\mathrm{F}_{\max }$ are also illustrated on the figure. If the advice for 2016-2017 fishing year, set according to FMSY, would be 21.5 kt as illustrated in figure 17.7.1

### 17.7.6 Future work

The model will need more work to be usable as basis for assessment. Main issues to explore at present is the length aggregation of the survey indices, for example it may be more prudent to have only one index for males and one for females rather than the 6 currently used for tuning the model. Work on aging is planned in Iceland-based on the Norwegian method for age determination. When this gets started and a few years of ageing data become available it is expected that the model will be more stable as was the case for tusk and ling in 5a (assessed in WGDEEP). Estimates of uncertainty were not presented based on the current model at the meeting due to time constraints.

### 17.8 Management Considerations

Available biological information and information on distribution of the fisheries suggest that Greenland halibut in East Greenland, Iceland and Faroe Islands belong to the same entity and do mix. Recent information of tagging experiments in the Barents Sea suggests high mixing between the Barents Sea and Iceland. This connectivity is not accommodated for in the present assessment.

In 2012 the coastal states initiated work on a common management plan for Greenland halibut. The aim was to have a common management plan to be implemented in 2015.

### 17.9 Data consideration and Assessment quality

The Icelandic cpue series has for many years been used as a biomass indicator in the assessment of the stock. The cpue of the Greenlandic trawlers and the biomass indices from the Faroese waters have not been used in the assessment, mainly because the stock production model is not able to accommodate contrasting indices (Icelandic cpue and Greenlandic/Icelandic autumn surveys).

### 17.10A number of issues on data and assessment quality are addressed in section 17.10.Proposals and recommendations

Stock structure and connectivity between the main fishing areas remains partly unknown. Basic biological information on spawning and nursery grounds for the juveniles also remains poorly known. Biomass indices over the entire assessment area are not similar to respect to trend over time and may suggest different dynamics between areas. Further, recent tagging experiments in the Barents Sea suggest a high connectivity with Iceland waters. Therefore a compilation of present knowledge of stock identification for Greenland halibut in the East Greenland, Iceland, Faroese and Norwegian waters should be made in order to review whether present stock areas are appropriate to assessment purposes. Such a compilation should be evaluated outside NWWG, e.g. by WGSIM.

A number of issues on the quality of the input biomass indices to the present assessment model are questioned. The Icelandic cpue series that is based on the principal trawler fleet is assumed to have undergone marked changes with respect to management regulations and spatial distribution. The possibility to estimate these effects by standardization of catch rates should be explored. Similar analyses should be conducted on the remaining cpue series, in order to evaluate them as indicative of biomass development.

The present assessment model, a stock production model in Bayesian framework, is criticized for its behavior in relation to the biomass indices. The models use of process error and sensitivity to various priors should be further scrutinized. A generic review of the model's performance could be by WGMG.

At the benchmark in 2013 (WKBUT) an alternative assessment model, Gadget, was presented. The group encouraged this model to be fully developed in order to replace the stock production model. Currently the Gadget model is not fully developed and several issues need further exploration (see section 17.7) and especially age data from the stock is required.

Ageing of Greenland halibut ceased for many of the marine institutes in Greenland, Iceland, Faroe Island and Norway around 2000 due to reading difficulties and lack of calibration. However, IMR in Norway have now developed a promising method to age

Greenland halibut and an ageing workshop is scheduled in August 2016 (WKAGE???). With the aim to revert to an age based assessment, it is suggested that cooperation between institutes is initiated and an inter calibration protocol is established. This task is a major task since a number of sampled otoliths back in time have to be read, and the time horizon for this project is therefore expected to exceed the near future. It is foreseen that the stock will be benchmarked in within the next years addressing the above issues.

### 17.11References

Sünksen,K. 2007. Bycatch in the fishery for Greenland halibut. WD 17, NWWG 2007.
Taylor L, Begley J, Kupca V, Stefansson G. 2007. A simple implementation of the statistical modelling framework Gadget for cod in Icelandic waters. African Journal of Marine Science 29: 223-245.

### 17.12Tables

Table 17.2.1 GREENLAND HALIBUT. Nominal landings (tonnes) by countries,
in Sub-areas V, VI, XII and XIV , as officially reported to ICES and estimated by WG


1) Provisional data
2) Iceland has no official reportings in 2015

Table 17.2.2 GREENLAND HALIBUT. Nominal landings (tonnes) by countries, in Division Va, as officially reported to ICES and estimated by WG.


Table 17.2.3 GREENLAND HALIBUT. Nominal landings (tonnes) by countries, in Division Vb as officially reported to ICES and estimated by WG.


1) Provisional data
2) WGestimate includes additional catches as described in Working Group reports for each year and in the report from 2001

Table 17.2.4 GREENLAND HALIBUT. Nominal landings (tonnes) by countries, in Sub-area XIV as officially reported to ICES and estimated by WG.

| Country | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | - | - | - | - | - | 78 | 74 | 98 | 87 |
| Germany | 2,893 | 2,439 | 1,054 | 818 | 636 | 745 | 456 | 595 | 420 |
| Greenland | + | 1 | 5 | 15 | 81 | 177 | 154 | 37 | 11 |
| Iceland | - | - | 1 | 2 | 36 | 17 | 136 | 40 | + |
| Norway | - | - | - | + | - | - | - | - | - |
| Russia | - | - | - | - | - | - | - | - | + |
| UK (Engl. and Wales) | - | - | - | - | - | - | - | - | - |
| UK (Scotland) | - | - | - | - | - | - | - | - | - |
| United Kingdom | - | - | - | - | - | - | - | - | - |
| Total | 2,893 | 2,440 | 1,060 | 835 | 753 | 1,017 | 820 | 770 | 518 |
| Working Group estimate | - | - | - | - | - | - | - | - | - |
| Country | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| Denmark | - | - | - | - | - | - | 1 | + | + |
| Faroe Islands | - | - | - | 181 | 168 | 147 | 130 | 148 | 151 |
| Germany | 293 | 279 | 311 | 391 | 639 | 808 | 3,343 | 3,301 | 3,399 |
| Greenland | 40 | 66 | 437 | 288 | 866 | 533 | 1,162 | 1,129 | $747^{1,7}$ |
| Iceland | - | - | - | 19 | 82 | 7 | - | 1,803 | 148 |
| Norway | 8 | 18 | 196 | 511 | 1,120 | 1,668 | 1,881 | 1,897 ${ }^{1}$ | 1,253 ${ }^{1}$ |
| Russia | - | - | 5 | - | - | 10 | 424 | 37 | 52 |
| UK (Engl. and Wales) | 27 | 38 | 108 | 796 | 513 | 1405 | 264 | 218 | 190 |
| UK (Scotland) | - | - | 18 | 26 | 84 | 205 | 13 |  |  |
| United Kingdom | - | - | - | - | - | - | - |  |  |
| Total | 368 | 401 | 1,075 | 2,212 | 3,472 | 4,783 | 7,218 | 8,533 | 5,940 |
| Working Group estimate | $736{ }^{2}$ | $875{ }^{3}$ | $1,176{ }^{4}$ | 2,249 ${ }^{5}$ | $3,125^{6}$ | 5,077 ${ }^{7}$ | 7,283 | 8,558 |  |
| Country | 1999 | 2000 | $2001{ }^{1}$ | $2002{ }^{1}$ | $2003{ }^{1}$ | $2004{ }^{1}$ | $2005{ }^{1}$ | $2006{ }^{1}$ | $2007{ }^{1}$ |
| Denmark |  |  |  |  |  |  |  |  |  |
| Faroe Islands | 2 |  |  | 274 | 366 | 274 | 186 | 22 |  |
| Germany | 3,047 | 3,243 | 2,750 | 2,019 | 2,925 | 5,159 | 5,144 | 4,298 | 4,702 |
| Greenland | $200{ }^{1,4}$ | 1,740 | 1,553 | 1,887 | 1,459 |  |  |  |  |
| Iceland | 93 | 30 | 14,280 | 16,947 | 6 |  |  |  |  |
| Ireland |  |  | 7 |  |  |  |  |  |  |
| Norway | 1,100 | 1,161 | 1,424 | 1,660 | 846 | 1,114 | 1,023 | 1,094 |  |
| Poland |  |  |  |  |  | 205 |  |  |  |
| Portugal |  |  | 6 | 130 |  |  |  | 1,094 |  |
| Russia | 138 | 183 | 186 | 44 |  | 261 |  | 505 | 500 |
| Spain |  | 8 | 10 |  | 2,131 | 3,406 | 2 |  |  |
| UK (Engl. and Wales) | 226 | 262 | 100 |  |  |  |  |  |  |
| UK (Scotland) |  |  |  | 24 | 188 | 278 | 160 |  |  |
| United Kingdom |  |  |  | 178 | 799 | 1,294 |  |  |  |
| Total | 4,806 | 6,627 | 20,316 | 22,889 ${ }^{\circ}$ | 8,720 | 11,991 | 6,515 | 7,013 | 5,202 |
| Working Group estimate | 0 | 6958 | $0^{6}$ | $0^{6}$ | 0 | 9,854 | 10,185 | 8,589 | 10,261 |
|  |  |  |  |  |  |  |  |  |  |
| Country | $2008{ }^{1}$ | $2009{ }^{1}$ | $2010{ }^{1}$ | $2011{ }^{\text {1 }}$ | $2012{ }^{1}$ | $2013{ }^{1}$ | $2014{ }^{1}$ | $2015{ }^{1}$ |  |
| Estonia |  |  |  |  |  |  | 429 |  |  |
| Faroe Islands |  | 270 | 333 |  | 77 | 125 | 409 | 57 |  |
| Germany | 4,842 | 4 | 4,490 | 5,206 | 4,351 | 3,428 | 3,114 | 3,543 |  |
| Greenland |  | 2,819 |  | 3,258 | 5,239 | 3,159 | 1,897 | 3,641 |  |
| Iceland |  |  |  |  | 7,290 |  | 3 | 46 |  |
| Ireland |  |  |  |  |  |  |  |  |  |
| Norway | 637 | 29 | 226 | 164 | 853 | 613 | 761 | 1,115 |  |
| Poland | 1,354 | 718 | 960 |  | 786 |  |  |  |  |
| Portugal |  |  |  |  |  |  |  |  |  |
| Russia | 763 |  | 1,070 | 1,095 | 1,168 | 1,369 | 587 | 600 |  |
| Spain |  |  |  |  |  |  |  |  |  |
| United Kingdom | 131 | 452 | 229 | 309 | 1 | 1 |  |  |  |
| Total | 7,727 | 4,292 | 7,308 | 10,032 | 19,765 | 8,694 | 7,200 | 9,002 |  |
| Working Group estimate | 0 | 9,805 | 10,402 | 10,761 |  |  | 7,526 | 9,534 |  |

1) Provisional data
2)WG estimate includes additional catches as described in working Group reports for each year and in the report from 2001.
2) Includes $125 t$ by Faroe Islands and $206 t$ by Greenland.
3) Excluding 4732 t reported as area unknown.
4) Includes 1523 t by Norway, 102 t by Faroe Islands, 3343 t by Germany, 1910 t by Greenland, 180 t by Russia, as reported to Greenland authorities.
5) Does not include most of the Icelandic catch as those are included in WG estimate of Va.
6) Excluding 138 t reported as area unknown.

Table 17.2.5 GREENLAND HALIBUT. Nominal landings (tonnes) by countries in Sub-area XII, as officially reported to the ICES and estimated by WG

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | $2003{ }^{1}$ | $2004{ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands |  | 47 |  |  |  |  | 40 |  |  |
| France |  |  |  |  | 1 |  |  | 4 | 30 |
| Ireland |  |  |  |  |  | 49 |  |  |  |
| Lithuania |  |  |  |  |  |  |  | 2 | 1 |
| Poland |  |  |  |  |  | 2 |  | 2 | 1 |
| Spain ${ }^{2}$ | 2 | 42 | 67 | 137 | 751 | 1338 | 28 | 730 | 1145 |
| UK |  |  |  |  | 7 | 5 |  |  |  |
| Russia |  |  |  |  |  |  |  |  |  |
| Norway | 2 |  |  |  | 553 | 500 | 316 | 201 | 119 |
| Estonia |  |  |  |  |  |  |  |  |  |
| Total | 4 | $89^{\prime \prime}$ | $67^{\circ}$ | 137 | 1,312 | 1,894 | 384 | 939 | 1,296 |
| WGestimate |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| Country | $2005{ }^{1}$ | $2006{ }^{1}$ | $2007{ }^{1}$ | $2008{ }^{1}$ | $2009{ }^{1}$ | $2010{ }^{1}$ | $2011{ }^{1}$ | $2012{ }^{1}$ | $2013{ }^{1}$ |
| Faroe Islands |  |  |  |  |  |  | 106 |  |  |
| France |  |  |  |  |  |  |  |  |  |
| Ireland |  |  |  |  |  |  |  |  |  |
| Lithuania |  | 2 | 3 | 566 |  |  |  | 97 |  |
| Poland |  |  |  |  |  |  |  |  |  |
| Spain ${ }^{2}$ | 501 |  |  |  |  |  |  |  |  |
| UK | 3 |  |  |  |  |  |  |  |  |
| Russia |  | 46 | 1 |  | 762 |  |  |  |  |
| Norway |  |  |  |  | 94 |  |  |  |  |
| Estonia |  | 2 |  |  |  |  |  |  |  |
| Total | 504 | 50 | 4 | 566 | 856 | 0 | 106 | 97 | 0 |
| WGestimate | 504 | 50 | 4 | 566 | 856 | 0 | 106 | 97 | 0 |



Table 17.2.6 GREENLAND HALIBUT. Nominal landings (tonnes) by countries in Sub-area VI, as officially reported to the ICES and estimated by WG.

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | $2003{ }^{1}$ | $2004{ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Estonia |  |  |  |  |  |  | 8 |  |  |
| Faroe Islands |  |  |  |  |  |  |  |  |  |
| France |  |  |  |  |  |  | 286 | 165 | 110 |
| Poland |  |  |  |  |  |  | 16 | 91 | 1 |
| Spain ${ }^{2}$ |  |  | 22 | 88 | 20 | 350 | 1367 | 214 | 170 |
| UK |  |  |  |  | 159 | 247 | 77 | 42 | 10 |
| Russia |  |  |  |  |  | 1 |  |  | 1 |
| Norway |  |  |  |  | 35 | 317 | 21 | 26 |  |
| Total | 0 | 0 | 22 | 88 | 214 | 915 | 1775 | 538 | 292 |
| WGestimate |  |  |  |  |  |  |  |  |  |
| Country | $2005{ }^{1}$ | $2006{ }^{1}$ | $2007{ }^{1}$ | $2008{ }^{1}$ | $2009{ }^{1}$ | $2010{ }^{1}$ | $2011{ }^{1}$ | $2012{ }^{1}$ | $2013{ }^{1}$ |
| Estonia | 5 | 1 |  |  |  |  |  |  |  |
| Faroe Islands |  |  |  |  |  | 1 |  |  | 0 |
| France |  | 22 | 8 | 114 |  | 38 | 8 | 54 | 113 |
| Poland |  |  |  |  |  |  |  |  |  |
| Spain ${ }^{2}$ | 3 | 33 |  |  |  |  |  |  |  |
| UK | 217 | 74 | 15 | 80 | 12 | 11 | 3 | 11 | 93 |
| Russia |  | 1 |  | 32 |  |  |  |  |  |
| Norway |  | 3 |  | 1 | 3 | 2 | 7 | 3 | 1 |
| Lithuania |  |  |  | 968 |  |  |  | 2 |  |
| Total | 225 | 134 | 23 | 1195 | 15 | 52 | 18 | 70 | 207 |
| WGestimate | 225 | 134 | 23 | 1195 | 15 | 52 | 18 | 70 | 207 |


| Country | $2014^{1}$ | $2015^{1}$ |
| :--- | ---: | ---: |
| Estonia |  |  |
| Faroe Islands | 1 |  |
| France |  | 89 |
| Poland |  |  |
| Spain 2 |  | 18 |
| UK | 42 | 119 |
| Russia | 0 | 1 |
| Norway <br> Lithuania | 43 | 227 |
| Total | 43 | 227 |
| WGestimate |  |  |

[^2]Table 17.3.1. Cpue indices of trawl fleets in Divisions 5a, 5b and 14b As Derived from Glm

| multiplicative models. |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | Year | Cpue | \% <br> Change <br> In Cpue <br> Between <br> Years | Landings | Rela- <br> tive <br> De- <br> rived <br> Effort | Relative <br> Derived <br> Effort | \% <br> Change <br> In Effort <br> Between <br> Years |
| Iceland 5a | 1985 | 1.00 |  | 29,197 | 29 | 100 |  |
|  | 1986 | 0.99 | -1 | 31,027 | 31 | 107 | 7 |
|  | 1987 | 0.96 | -3 | 44,659 | 47 | 149 | 38 |
|  | 1988 | 0.91 | -5 | 49,379 | 54 | 117 | -21 |
|  | 1989 | 0.79 | -13 | 59,272 | 75 | 138 | 18 |
|  | 1990 | 0.75 | -5 | 37,308 | 50 | 66 | -52 |
|  | 1991 | 0.73 | -2 | 35,413 | 48 | 97 | 47 |
|  | 1992 | 0.67 | -9 | 31,978 | 48 | 100 | 2 |
|  | 1993 | 0.53 | -20 | 34,134 | 64 | 133 | 33 |
|  | 1994 | 0.44 | -18 | 28,608 | 65 | 102 | -23 |
|  | 1995 | 0.36 | -19 | 27,391 | 77 | 118 | 15 |
|  | 1996 | 0.30 | -15 | 22,073 | 73 | 94 | -20 |
|  | 1997 | 0.32 | 5 | 16,792 | 53 | 72 | -24 |
|  | 1998 | 0.50 | 57 | 10,595 | 21 | 40 | -44 |
|  | 1999 | 0.55 | 9 | 11,138 | 20 | 96 | 139 |
|  | 2000 | 0.59 | 7 | 14,607 | 25 | 122 | 27 |
|  | 2001 | 0.59 | 1 | 16,752 | 28 | 113 | -8 |
|  | 2002 | 0.48 | -20 | 19,714 | 41 | 147 | 29 |
|  | 2003 | 0.35 | -26 | 20,415 | 58 | 139 | -5 |
|  | 2004 | 0.30 | -16 | 15,477 | 52 | 90 | -35 |
|  | 2005 | 0.27 | -8 | 13,172 | 48 | 92 | 2 |
|  | 2006 | 0.37 | 33 | 11,817 | 32 | 67 | -27 |
|  | 2007 | 0.46 | 25 | 10,525 | 23 | 71 | 6 |
|  | 2008 | 0.40 | -13 | 9,580 | 24 | 105 | 48 |
|  | 2009 | 0.41 | 4 | 15,782 | 38 | 158 | 50 |
|  | 2010 | 0.41 | -1 | 13,565 | 33 | 87 | -45 |
|  | 2011 | 0.43 | 4 | 14,048 | 33 | 99 | 15 |
|  | 2012 | 0.44 | 3 | 7,312 | 17 | 51 | -49 |
|  | 2013 | 0.45 | 2 | 15,439 | 35 | 208 | 309 |
|  | 2014 | 0.41 | -7 | 10,475 | 25 | 73 | -65 |
|  | 2015 | 0.45 | 8 | 12,593 | 28 | 112 | 53 |
| Greenland, 14b | 1991 | 1.00 |  | 875 | 1 | 100 | 0 |
|  | 1992 | 0.92 | -8 | 1,176 | 1 | 147 | 47 |
|  | 1993 | 2.48 | 171 | 2,249 | 1 | 71 | -52 |
|  | 1994 | 3.19 | 29 | 3,125 | 1 | 108 | 53 |
|  | 1995 | 3.26 | 2 | 5,077 | 2 | 159 | 47 |
|  | 1996 | 3.24 | -1 | 7,283 | 2 | 144 | -9 |
|  | 1997 | 3.35 | 3 | 8,558 | 3 | 114 | -21 |


| MULTIPLICATIVE MODELS. |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | Year | Cpue | \% <br> Change <br> In Cpue <br> Between <br> Years | Landings | Relative Derived Effort | Relative Derived Effort | \% <br> Change <br> In Effort <br> Between <br> Years |
|  | 1998 | 3.28 | -2 | 5,940 | 2 | 71 | -38 |
|  | 1999 | 2.29 | -30 | 5,376 | 2 | 130 | 83 |
|  | 2000 | 2.11 | -8 | 6,958 | 3 | 141 | 9 |
|  | 2001 | 2.19 | 4 | 7,216 | 3 | 100 | -29 |
|  | 2002 | 2.38 | 9 | 6,621 | 3 | 85 | -15 |
|  | 2003 | 2.33 | -2 | 8,017 | 3 | 124 | 46 |
|  | 2004 | 2.27 | -2 | 9,854 | 4 | 126 | 2 |
|  | 2005 | 3.14 | 38 | 10,185 | 3 | 75 | -41 |
|  | 2006 | 3.24 | 3 | 8590 | 3 | 82 | 9 |
|  | 2007 | 3.07 | -5 | 10261 | 3 | 126 | 54 |
|  | 2008 | 3.11 | 1 | 8,952 | 3 | 86 | -32 |
|  | 2009 | 2.58 | -17 | 10,567 | 4 | 143 | 66 |
|  | 2010 | 2.70 | 5 | 10,402 | 4 | 94 | -34 |
|  | 2011 | 2.66 | -2 | 10,761 | 4 | 105 | 12 |
|  | 2012 | 3.14 | 18 | 12,475 | 4 | 98 | -7 |
|  | 2013 | 2.91 | -7 | 12,476 | 4 | 108 | 10 |
|  | 2014 | 3.09 | 6 | 7,526 | 2 | 57 | -47 |
|  | 2015 | 3.50 | 13 | 9,534 | 3 | 112 | 97 |
| Faroe Islands, 5b | 1991 | 1.00 |  | 1,662 | 2 | 100 | 34 |
|  | 1992 | 0.34 | -21 | 2,269 | 7 | 397 | 297 |
|  | 1993 | 0.24 | -11 | 4,434 | 19 | 282 | -29 |
|  | 1994 | 0.23 | -2 | 5,225 | 23 | 121 | -57 |
|  | 1995 | 0.16 | -28 | 3,832 | 23 | 103 | -15 |
|  | 1996 | 0.17 | 4 | 6,469 | 37 | 160 | 55 |
|  | 1997 | 0.19 | 12 | 4,870 | 25 | 67 | -58 |
|  | 1998 | 0.14 | -34 | 3,825 | 28 | 112 | 67 |
|  | 1999 | 0.16 | 12 | 4,265 | 27 | 96 | -15 |
|  | 2000 | 0.17 | 11 | 5,079 | 29 | 109 | 14 |
|  | 2001 | 0.20 | 19 | 3,245 | 16 | 55 | -50 |
|  | 2002 | 0.16 | -24 | 2,694 | 17 | 104 | 91 |
|  | 2003 | 0.10 | -29 | 2,426 | 24 | 141 | 35 |
|  | 2004 | 0.08 | -12 | 1,771 | 21 | 89 | -37 |
|  | 2005 | 0.09 | 4 | 892 | 10 | 48 | -46 |
|  | 2006 | 0.10 | 19 | 873 | 8 | 83 | 72 |
|  | 2007 | 0.12 | 16 | 1,060 | 9 | 107 | 28 |
|  | 2008 | 0.18 | 60 | 1735 | 10 | 107 | 0 |
|  | 2009 | 0.21 | 26 | 1760 | 8 | 87 | -19 |
|  | 2010 | 0.17 | -21 | 1,413 | 8 | 98 | 13 |
|  | 2011 | 0.31 | 65 | 1,489 | 5 | 59 | -40 |
|  | 2012 | 0.30 | -4 | 2,163 | 7 | 148 | 153 |


| Area | Year | Cpue | \% <br> Change <br> In Cpue <br> Between <br> Years | Landings | Relative Derived Effort | Relative Derived Effort | \% <br> Change <br> In Effort Between Years |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2013 | NA |  | 2,560 |  |  |  |
|  | 2014 | NA |  | 2,958 |  |  |  |
|  | 2015 | NA |  | 3,139 |  |  |  |

Table 17.6.1. Assessment input dataseries: Catch by the fishery; three indices of stock biomass - a standardized catch rate index based on fishery data (cpue) from the Iceland EEZ, a Icelandic (Ice) and a Greenlandic (Green) research survey index.

| Year | Catch (ktons) | $\begin{aligned} & \hline \text { CPUE } \\ & \text { (index) } \\ & \hline \end{aligned}$ | Survey (ktons) |
| :---: | :---: | :---: | :---: |
| 1960 | 0 | - | - |
| 1961 | 0.029 | - | - |
| 1962 | 3.071 | - | - |
| 1963 | 4.275 | - | - |
| 1964 | 4.748 | - | - |
| 1965 | 7.421 | - | - |
| 1966 | 8.030 | - | - |
| 1967 | 9.597 | - | - |
| 1968 | 8.337 | - | - |
| 1969 | 26.200 | - | - |
| 1970 | 33.823 | - | - |
| 1971 | 28.973 | - | - |
| 1972 | 26.473 | - | - |
| 1973 | 20.463 | - | - |
| 1974 | 36.280 | - | - |
| 1975 | 23.494 | - | - |
| 1976 | 6.045 | - | - |
| 1977 | 16.578 | - | - |
| 1978 | 14.349 | - | - |
| 1979 | 23.622 | - | - |
| 1980 | 31.157 | - | - |
| 1981 | 19.239 | - | - |
| 1982 | 32.441 | - | - |
| 1983 | 30.891 | - | - |
| 1984 | 34.024 | - | - |
| 1985 | 32.075 | 1.76 | - |
| 1986 | 32.984 | 1.74 | - |
| 1987 | 46.622 | 1.69 | - |
| 1988 | 51.118 | 1.60 | - |
| 1989 | 61.396 | 1.85 | - |
| 1990 | 39.326 | 1.32 | - |
| 1991 | 37.950 | 1.29 | - |
| 1992 | 35.487 | 1.17 | - |
| 1993 | 41.247 | 0.94 | - |
| 1994 | 37.190 | 0.77 | - |
| 1995 | 36.288 | 0.63 | - |
| 1996 | 35.932 | 0.53 | 66 |
| 1997 | 30.309 | 0.56 | 90 |
| 1998 | 20.382 | 0.88 | 91 |
| 1999 | 20.371 | 0.96 | 90 |
| 2000 | 26.644 | 1.03 | 101 |
| 2001 | 27.291 | 1.04 | 110 |
| 2002 | 29.158 | 0.84 | 84 |
| 2003 | 30.891 | 0.62 | 52 |
| 2004 | 27.102 | 0.52 | 36 |
| 2005 | 24.249 | 0.48 | 56 |
| 2006 | 21.432 | 0.64 | 39 |
| 2007 | 20.957 | 0.80 | 50 |
| 2008 | 22.169 | 0.70 | 58 |
| 2009 | 27.349 | 0.73 | 80 |
| 2010 | 25.995 | 0.72 | 59 |
| 2011 | 26.424 | 0.75 | 71 |
| 2012 | 29.309 | 0.77 | 82 |
| 2013 | 27.045 | 0.79 | 85 |
| 2014 | 21.069 | 0.73 | 75 |
| 2015 | 25.677 | 0.78 | 80 |
| 2016* | 25.000 |  |  |

Table 17.6.2. Priors used in the assessment model. ~ means "distributed as..", dunif = uniform-, dlnorm $=$ lognormal-, dnorm= normal- and dgamma $=$ gammadistributed. Symbols as in text.

| Parameter |  | Prior |  |
| :---: | :---: | :---: | :---: |
| Name | Symbol | Type | Distribution |
| Maximal Suatainable Yield | MSY | reference | dunif $(1,300)$ |
| Carrying capacity | $K$ | low informative | dnorm( 750,300 ) |
| Catchability Iceland survey | $q_{\text {lce }}$ | reference | $\ln \left(\mathrm{q}_{\text {ice }}\right) \sim$ dunif( $\left.-3,1\right)$ |
| Catchability Greenland survey | $q_{\text {Green }}$ | reference | $\ln \left(\mathrm{q}_{\text {Green }}\right) \sim$ dunif( $-3,1$ ) |
| Catchability Iceland CPUE | $q_{\text {cpue }}$ | reference | $\ln \left(\mathrm{q}_{\text {cpue }}\right) \sim$ dunif( $-10,1$ ) |
| Initial biomass ratio | $P_{1}$ | informative | dnorm( $2,0.071$ ) |
| Precision Iceland survey | $1 / \sigma_{I C e}{ }^{2}$ | low informative | dgamma(2.5,0.03) |
| Precision Greenland survey | $1 / \sigma_{\text {Green }}{ }^{2}$ | low informative | dgamma(2.5,0.03) |
| Precision Iceland CPUE | $1 / \sigma_{\text {cpue }}{ }^{2}$ | low informative | dgamma(2.5,0.03) |
| Precision model | $1 / \sigma_{P}{ }^{2}$ | reference | dgamma(0.01, 0.01 ) |

Table 17.6.3. Summary of parameter estimates: mean, standard deviation (sd) and 25, 50, and 75 percentiles of the posterior distribution of selected parameters (symbols as in the text).

|  | Mean | sd | $25 \%$ | Median | $75 \%$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
| $M S Y$ (ktons) | 33.71 | 11.82 | 26.63 | 33.01 | 39.81 |
| $K$ (ktons) | 893 | 272 | 701 | 878 | 1072 |
| $r$ | 0.17 | 0.08 | 0.11 | 0.16 | 0.21 |
| $q_{\text {cpue }}$ | 0.003 | 0.001 | 0.002 | 0.003 | 0.003 |
| $q_{\text {Survey }}$ | 0.27 | 0.12 | 0.19 | 0.24 | 0.31 |
| $P_{1985}$ | 1.57 | 0.13 | 1.48 | 1.57 | 1.66 |
| $P_{\text {2016 }}$ | 0.72 | 0.11 | 0.65 | 0.71 | 0.79 |
| $\sigma_{\text {cpue }}$ | 0.10 | 0.02 | 0.08 | 0.09 | 0.11 |
| $\sigma_{\text {Survey }}$ | 0.18 | 0.03 | 0.16 | 0.18 | 0.20 |
| $\sigma_{P}$ | 0.16 | 0.03 | 0.14 | 0.16 | 0.18 |

Table 17.6.4. Model diagnostics: residuals (\% of observed value), probability of getting a more extreme observation (p.extreame; see text for explanation).

|  | CPUE |  |  | Survey |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Year | resid $(\%)$ | Pr | resid $(\%)$ | Pr |
| 1985 | -1.73 | 0.55 | - | - |
| 1986 | -0.67 | 0.52 | - | - |
| 1987 | 0.31 | 0.49 | - | - |
| 1988 | 2.43 | 0.42 | - | - |
| 1989 | -8.26 | 0.75 | - | - |
| 1990 | 3.38 | 0.39 | - | - |
| 1991 | -1.53 | 0.55 | - | - |
| 1992 | -2.95 | 0.59 | - | - |
| 1993 | 0.21 | 0.49 | - | - |
| 1994 | 0.74 | 0.48 | - | - |
| 1995 | 3.81 | 0.38 | - | - |
| 1996 | 12.17 | 0.16 | -14.16 | 0.76 |
| 1997 | 16.11 | 0.10 | -35.72 | 0.96 |
| 1998 | -3.39 | 0.62 | -11.09 | 0.71 |
| 1999 | -1.43 | 0.55 | 0.62 | 0.49 |
| 2000 | -1.26 | 0.54 | -3.68 | 0.58 |
| 2001 | -2.95 | 0.60 | -12.93 | 0.75 |
| 2002 | -1.51 | 0.55 | -5.90 | 0.62 |
| 2003 | -0.04 | 0.50 | 13.18 | 0.25 |
| 2004 | -1.59 | 0.55 | 30.79 | 0.06 |
| 2005 | 7.34 | 0.27 | -12.45 | 0.74 |
| 2006 | -9.44 | 0.78 | 35.77 | 0.04 |
| 2007 | -14.73 | 0.88 | 27.99 | 0.08 |
| 2008 | -1.42 | 0.54 | 12.96 | 0.25 |
| 2009 | 0.64 | 0.48 | -12.90 | 0.75 |
| 2010 | -1.33 | 0.54 | 14.24 | 0.23 |
| 2011 | -0.30 | 0.51 | 0.87 | 0.48 |
| 2012 | 1.86 | 0.44 | -8.90 | 0.68 |
| 2013 | 0.70 | 0.48 | -10.88 | 0.71 |
| 2014 | 3.66 | 0.38 | -3.37 | 0.57 |
| 2015 | 1.33 | 0.46 | -5.62 | 0.61 |
|  |  |  |  |  |

Table 17.6.5. Upper: stock status for 2015 and predicted to the end of 2016. Lower: predictions for 2017 with catch options from $0-30$ ktons and the catch option corresponding to Fmš.

| Status | 2015 | 2016 * |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Risk of falling below $B_{\text {msy_trigger }}$ | 0\% | 0\% |  |  |  |  |  |
| Risk of falling below $B_{M S Y}$ | 100\% | 93\% |  |  |  |  |  |
| Risk of exceeding $F_{M S Y}$ | 43\% | 56\% |  |  |  |  |  |
| Risk of exceeding $F_{\text {lim }}\left(1.7 F_{M S Y}\right)$ | 9\% | 15\% |  |  |  |  |  |
| Stock size (B/Bmsy), median | 0.71 | 0.73 |  |  |  |  |  |
| Fishing mortality (F/Fmsy), | 1.10 | 1.06 |  |  |  |  |  |
| Productivity (\% of MSY) | 92\% | 92\% |  |  |  |  |  |
| *Predicted catch in $2016=25 \mathrm{ktons}$ |  |  |  |  |  |  |  |
| Catch option 2017 (ktons) | 0 | 5 | 10 | 15 | 20 | 24 | 30 |
| Prob. of falling below $B_{\text {LIM }}$ | 0\% | 0\% | 0\% | 0\% | 0\% | 0\% | 0\% |
| Risk of falling below $B_{M S Y}$ | 81\% | 82\% | 84\% | 85\% | 86\% | 86\% | 88\% |
| Risk of exceeding $F_{M S Y}$ |  | 1\% | 5\% | 16\% | 34\% | 50\% | 71\% |
| Risk of exceeding $F_{\text {lim }}\left(1.7 F_{M S Y}\right)$ | - | 0\% | 2\% | 4\% | 9\% | 15\% | 27\% |
| Stock size (B/Bmsy), median | 0.80 | 0.78 | 0.77 | 0.76 | 0.75 | 0.75 | 0.73 |
| Fishing mortality (F/Fmsy), | - | 0.20 | 0.40 | 0.61 | 0.82 | 0.99 | 1.27 |
| Productivity (\% of MSY) | 96\% | 95\% | 95\% | 94\% | 94\% | 94\% | 93\% |

### 17.13Figures



Figure 17.2.1. Landings of Greenland halibut in Divisions 5, 12 and 14. As the landings within Icelandic waters, since 1976, have not officially been separated and reported according to the defined ICES statistical areas, they are set under area 5a by the North Western Working Group. In 2012 Icelandic landings in 14 were partly recorded in 14, while for remaining years all landings are recorded in 5 a.


Figure 17.2.2 Greenland halibut 5+14. Distribution of fishing effort in 2015.500 m and 1000 m depth contours are shown.


Figure 17.2.3. Greenland halibut 5+14. Distribution of catches in the fishery in 2015.500 m and 1000 $m$ depth contours are shown.


Figure 17.2.4. Greenland halibut 5+14. Distribution of total fishing effort 2000-2015. The 500 m and 1000 m depth contours are shown.


Figure 17.2.5. Greenland halibut 5+14. Distribution of total catches in the fishery 2000-2015 500m and 1000 m depth contours are shown.


Figure 17.3.1. Standardized cpues from the Icelandic trawler fleet in Va. Area 1-4 are west, north, east and southeast. The average index of the four areas are used as biomass indicator in the stock production model.


Figure 17.3.2 Standardized cpue from the Icelandic trawler fleet in 5a by four main fishing areas in Va. 95\% CI indicated.


Figure 17.3.3. Standardized cpue, observed and derived effort from Icelandic trawl fishery.


Figure 17. 3.4. Standardized cpue from the Faroese trawler fleet. 95\% CI indicated


Figure 17.3.5. Standardized cpue from trawler fleets in 14b. $95 \%$ CI indicated. Points are raw observations.


Figure 17.3.6. Standardized cpue from trawler fleets in $14 b$ shown by subdivisions in $14 b$ in a north-south direction. $95 \%$ CI indicated.


Figure 17.4.1. Length distributions from the commercial trawl fishery in the western fishing grounds of Iceland (5a) in the years 2002-2015. Blue indicate males and red indicates females.


Figure 17.4.2. Length distributions from the commercial fishery in Subareas 5 and 14 by gear (BMT=bottom trawl, LLN=longlines, SHT=shrimp trawl).


Figure 17.5.1. Stations covered by scientific surveys in $14+5$ indicated as station positions in 2015 by the Greenland ( $n=76$ ) and Iceland ( $n=203$ ).


Figure 17.5.2. Distribution of Greenland halibut catch rates from the combined Greenland-Icelandic fall survey since 1996.


Figure 17.5.3. Greenland halibut in Icelandic fall groundfish survey. No survey was conducted in 2011.


Figure 17.5.4. Abundance indices by length for the Icelandic fall survey 1996-2015. No survey was conducted in 2011.


Figure 17.5.5. Catch rates from a combined survey/fisher's survey in 5b. Estimates are from a GLM model.


Figure 17.5.6. Distribution of catches of Greenland halibut at East Greenland in 1998-2009 in the Greenland deep-water survey.


Figure 17.5.6 continued. Distribution of catches of Greenland halibut at East Greenland in 2010 2015 in the Greenland deep-water survey.


Figure 17.5.7. Standardized catch rates from the Greenland survey.(95\% CI indicated.)


Figure 17.6.1. Probability density distributions of model parameters: estimated posterior (solid line) and prior (broken line) distributions


Figure 17.6.2. Observed (red curve) and predicted (dashed lines) series of the two biomass indices input to the model. Dashed lines are inter-quartile range of the posteriors.


Figure 17.6.3. Retrospective plot of median relative biomass $\left(B / B_{m s y}\right)$.


Figure 17.6.4. Stock trajectory. Estimated annual median biomass-ratio ( $\mathrm{B}^{\mathbf{~}} \mathrm{B}_{\mathrm{MSY}}$ ) and fishing mortal-ity-ratio ( $\mathbf{F} / \mathrm{F}_{\mathrm{MSY}}$ ). $\mathrm{B}_{\text {lim, }}$ MSY $\mathrm{B}_{\text {trigger }}$ and $\mathrm{F}_{\text {lim }}$ are indicated.


Figure 17.6.5. Stock summary, upper panel right: fishing mortality ( $\mathrm{F} / \mathrm{Fmš}_{\text {亿 }}$ ) and $95 \%$ conf limits, left: total biomass ( $\mathrm{B} / \mathrm{B}_{\mathrm{ms}}$ ) and $95 \%$ conf limits and lower panel is landings since start of the fishery. MSY $B_{\text {trigger }}$ (green dashed line), $\mathrm{B}_{\text {lim }}$ and $\mathrm{Flim}_{\text {( }}$ (blue dashed lines) are indicated.


Figure 17.6.6 Estimated time-series of relative biomass ( $B_{t} / B_{m s y}$ ) under different catch option scenarios: $0,5,10,15$ and 24 kt from upper to lower panel. Bold red lines are inter-quartile ranges and the solid black line is the median; the error bars extend to cover the central 90 per cent of the distribution.


Figure 17.6.7. Projections: Medians of estimated posterior biomass- and fishing mortality ratios; estimated risk of exceeding $F_{M S Y}$ or going below and $B_{M S Y t r i g g e r}$ given catch ranges at 0-30 ktons.


Figure 17.6.8. Historic landings and projected landings 2015-2025 under various $F$ ratio options from 0.3-1.7 F/Fmsy Solid red line is median, quartiles and $90 \%$ conf limit indicated.


Figure 17.6.9. The logistic production curve in relation to stock biomass ( $\mathbf{B} / \mathrm{B}_{\mathrm{ms}}$ ) (upper) and fishing mortality ( $\mathrm{F} / \mathrm{F}_{\mathrm{MSY}}$ ) (lower). Upper: points of maximum sustainable yield (MSY) and corresponding stock size are shown as well as the slope (red line) of the production curve (blue line); lower: points of MSY and corresponding fishing mortality and $\mathrm{Fcrash}\left(\mathrm{F} \geq \mathrm{F}_{\text {crash }}\right.$ do not have stable equilibriums and will drive the stock to zero).


Figure 17.7.1: Population estimates from the exploratory Gadget assessment model. Dashed horizontal line indicates start of the deterministic projection period based on different values of $F$, i.e. $F_{0.1}, F_{p a}, F_{\text {msy }}$ and $F_{\text {max }}$.


Figure 17.7.2 Estimates of the equilibrium catch curves and spawning-stock biomass by harvest rate (black solid line) with the associate $95 \%$ confidence regions (yellow). The red region illustrates the $B_{\text {lim. }}$. The different F targets are illustrated with vertical lines.

## 18 Redfish in Subareas 5, 6, 12 and 14

This chapter deals with fisheries directed to Sebastes species in Subareas 5, 6, 12 and 14 (chapters 18.4 and 18.7), and the abundance and distribution of juveniles (chapter 18.2.1), among other issues.

The "Workshop on Redfish Stock Structure" (WKREDS, 22-23 January 2009, Copenhagen, Denmark; ICES 2009) reviewed the stock structure of Sebastes mentella in the Irminger Sea and adjacent waters. ACOM concluded, based on the outcome of the WKREDS meeting, that there are three biological stocks of $S$. mentella in the Irminger Sea and adjacent waters:

- a 'Deep Pelagic' stock (NAFO 1-2, ICES V, $12,14>500 \mathrm{~m}$ ) - primarily pelagic habitats, and including demersal habitats west of the Faroe Islands;
- a 'Shallow Pelagic' stock (NAFO 1-2, ICES V, 12, $14<500 \mathrm{~m}$ ) - extends to ICES 1 and 2 , but primarily pelagic habitats, and includes demersal habitats east of the Faroe Islands;
- an ‘Icelandic Slope' stock (ICES 5a, 14)-primarily demersal habitats.

This conclusion is primarily based on genetic information, i.e. microsatellite information, and supported by analysis of allozymes, fatty acids and other biological information on stock structure, such as some parasite patterns. The Russian Federation maintains the point of view that there is only one stock of S. mentella in the pelagic waters of the Irminger Sea. Accordingly, the Russian Federation presented alternative approaches to stock assessment as well as environmental influence on stock dynamics. Briefly, it is claimed that the current survey based assessment does not adequately reflect stock status and that environmental factors - temperature causes major distributional changes of redfish-affect stock status more than fisheries and the use of the current management areas is rejected (see WD27 and Annex 7). The other NWWG members did not agree with the Russian Federation's view on stock structure and did not consider the presented assessment approach sufficiently documented.

The adult redfish on the Greenland shelf has traditionally been attributed to several stocks, and there remains the need to investigate the affinity of adult S. mentella in this region. The East-Greenland shelf is most likely a common nursery area for the three biological stocks.

ICES past advice for $S$. mentella fisheries was provided for two distinct management units, i.e. a demersal unit on the continental shelves and slopes and pelagic unit in the Irminger Sea and adjacent waters. However, based on the new stock identification information, ICES recommends three potential management units that are geographic proxies for biological stocks that were partly defined by depth and whose boundaries are based on the spatial distribution pattern of the fishery to minimize mixed-stock catches (see Figure 18.1.1):

- Management Unit in the northeast Irminger Sea: ICES Areas 5a, 12, and 14.
- Management Unit in the southwest Irminger Sea: NAFO Areas 1 and 2, ICES areas 5b, 12 and 14.
- Management Unit on the Icelandic slope: ICES Areas 5a and 14, and to the north and east of the boundary proposed in the $M U$ in the northeast Irminger Sea.
The pelagic fishery in the Irminger Sea and adjacent waters shows a clear distinction between two widely separated grounds fished at different seasons and depths. Spatial
analysis of the pelagic fishery catch and effort by depth, inside and outside the boundaries proposed for the management units in the northeast Irminger Sea, indicate that the boundaries effectively delineate the pelagic fishery in the northeast Irminger Sea from the pelagic fishery in the southwest Irminger Sea, with a small portion of mixedstock catches. In the last decade the majority (more than $90 \%$ ) of the catches have been taken in the northeast Irminger Sea. The northeastern fisheries on the pelagic S. mentella occur at the start of the fishing season at depths below 500 m and overlap to some extent with demersal fisheries on the continental slopes of Iceland (Sigurdsson et al., 2006).

A schematic illustration of the relationship between the management units and biological stocks is given in Figure 18.1.2.

For the abovementioned reasons, the Group now provides advice for the following Sebastes units:

- the S. norvegicus on the continental shelves of ICES Divisions 5a, 5b and Subarea 6 and 14 (chapter 19);
- the demersal S. mentella on the Icelandic slope (chapter 20);
- the shallow and deep pelagic S. mentella units in the Irminger Sea and adjacent waters (chapters 21 and 22 , respectively)
- the Greenland shelf S. mentella (chapter 23).


### 18.1 Environmental and ecosystem information

Species of the genus Sebastes are common and widely distributed in the North Atlantic. They are found off the coast of Great Britain, along Norway and Spitsbergen, in the Barents Sea, off the Faroe Islands, Iceland, East and West Greenland, and along the east coast of North America from Baffin Island to Cape Cod. All Sebastes species are viviparous. Copulation occurs in autumn-early winter and larvae extrusion takes place in late winter-late spring/early summer. Little is known about the copulation areas.

The increase of water temperature in the Irminger Sea may have an effect on spatial and vertical distribution of S. mentella in the feeding area (Pedchenko, 2005). The abundance and distribution of pelagic $S$. mentella in relation to oceanographic conditions were analysed in a special multistage workshop (ICES 2012). Based on 20 years of survey data, the results reveal the average relation of pelagic redfish to their physical habitat in shallow and intermediate waters: The most preferred latitude, longitude, depth, salinity and temperature for $S$. mentella are approximately $58^{\circ} \mathrm{N}, 40^{\circ} \mathrm{W}, 300 \mathrm{~m}, 34.89$ and $4.4^{\circ} \mathrm{C}$, respectively. The spatial distribution of $S$. mentella in the Irminger Sea mainly in waters $<500 \mathrm{~m}$ (and thus mainly relating to the "shallow" stock) appears strongly influenced by the Irminger Current Water (ICW) temperature changes, linked to the Subpolar Gyre (SPG) circulation and the North Atlantic Oscillation (NAO). The fish avoid waters mainly associated with the ICW ( $>4.5^{\circ} \mathrm{C}$ and $>34.94$ ) in the northeastern Irminger Sea, which may cause displacement of the fish towards the southwest, where fresher and colder water occurs.

Results based on international redfish survey data suggest that the interannual distribution of fish above 500 m will shift in a southwest/northeast direction depending on integrated oceanographic conditions (ICES 2012).

### 18.2 Environmental drivers of productivity

### 18.2.1 Abundance and distribution of 0 -group and juvenile redfish

Available data on the distribution of juvenile S. marinus indicate that the nursery grounds are located in Icelandic and Greenland waters. No nursery grounds have been found in Faroese waters. Studies indicate that considerable amounts of juvenile $S$. norvegicus off East Greenland are mixed with juvenile S. mentella (Magnússon et al. 1988; 1990, ICES CM 1998/G:3). The 1983 Redfish Study Group report (ICES CM 1983/G:3) and Magnússon and Jóhannesson (1997) describe the distribution of 0-group S. norvegicus off East Greenland. The nursery areas for S. norvegicus in Icelandic waters are found all around Iceland, but are mainly located west and north of the island at depths between 50 and 350 m (ICES CM 1983/G:3; Einarsson, 1960; Magnússon and Magnússon 1975; Pálsson et al. 1997). As they grow, the juveniles migrate along the north coast towards the most important fishing areas off the west coast.

Indices for 0-group redfish in the Irminger Sea and at East Greenland areas were available from the Icelandic 0-group surveys from 1970-1995. Thereafter, the survey was discontinued. Above average year-class strengths were observed in 1972, 1973-1974, 1985-1991, and in 1995.

There are very few juvenile demersal S. mentella in Icelandic waters (see chapter 20), and the main nursery area for this species is located off East Greenland (Magnússon et al. 1988, Saborido-Rey et al. 2004). Abundance and biomass indices of redfish smaller than 17 cm from the German annual groundfish survey, conducted on the continental shelf and slope of West and East Greenland down to 400 m , show that juveniles were abundant in 1993 and 1995-1998 (Figure 18.2.1). The 1999-2006 survey results indicate low abundance and were similar to those observed in the late 1980s. Since 2008, the survey index has been very low and was in 2013-2015 the lowest value recorded since 1982. Juvenile redfish were only classified to the genus Sebastes spp., as identification of small specimens to species level is difficult due to very similar morphological features. Observations on length distributions of S. mentella fished deeper than 400 m indicate that a part of the juvenile S. mentella on the East Greenland shelf migrates into deeper shelf areas and into the pelagic zone in the Irminger Sea and adjacent waters (Stransky 2000), with unknown shares.

### 18.3 Ecosystem considerations

Information on the ecosystems around the Faroe Islands, Iceland and Greenland is given in chapters 2, 7 and 13.

Analysis of the oceanographic situation in the Irminger Sea during the 2013 international survey and long-term data including 2003, allows the following conclusions:

Strong positive anomalies of temperature observed in the upper layer of the Irminger Sea with a maximum in 1998 are related to an overall warming of water in the Irminger Sea and adjacent areas in 1994-2013. These changes were also observed in the Irminger Current above the Reykjanes Ridge (Pedchenko, 2000), off Iceland (Malmberget al., 2001) and in the Labrador Seawater (Mortensen and Valdimarsson, 1999). Thus, temperature and salinity in the Irminger Current have increased since 1997 to the highest values seen for decades (ICES, 2001).

The 2003 survey detected high temperature anomalies within the $0-200 \mathrm{~m}$ layer in the Irminger Sea and adjacent waters. At 200-500 m depth and deeper waters, positive
anomalies were observed in most of the surveyed area. However, increasing temperature as compared to the survey in June-July 2001 was detected only north of $60^{\circ} \mathrm{N}$ in the flow of the Irminger Current above the Reykjanes Ridge and the northwestern part of the Irminger Sea. These changes in oceanographic conditions might have an effect on the seasonal distribution of redfish and its aggregations in the layer shallower than 500 m in the survey area (ICES, 2003).
In June/July 2005 and 2007, water temperature in the shallower layer ( $0-500 \mathrm{~m}$ ) of the Irminger Sea was higher than normal (ICES, 2005). As in the surveys 1999-2003, the redfish were aggregating in the southwestern part of the survey area, partly influenced by these hydrographic conditions. Favorable conditions for aggregation of redfish in an acoustic layer have been marked only in the southwestern part of the survey area with temperatures between $3.6-4.5^{\circ} \mathrm{C}$, as confirmed by the survey results obtained in 2009.

The hydrography in the survey of June/July 2013 shows that temperature in the survey area is above average but it was lower than in 2011 in most of the surveyed area, except for the Irminger Current (ICES, 2013).

### 18.4 Description of fisheries

There are three species of commercially exploited redfish in ICES Subarea 5, 6, 12, and 14: S. norvegicus (in publication both names S. norvegicus and S. marinus can be found, but according to Fernholm and Wheeler (1983) the first name is the correct name), $S$. mentella and S. viviparus. S. viviparus has only been of a minor commercial value in Icelandic waters and it is exploited in two small areas south of Iceland at depths of 150250 m . The landings of S. viviparus decreased from 1160 t in 1997 to 2-9 t in 2003-2006 (Table 18.4.1) due to decreased commercial interest in this species. The landings in 2009 amounted to 37 t , more than a twofold increase compared with 2008. After a directed fishery developed in 2010, with a total catch of 2600 t , the MRI advised on a 1500 t TAC for the 2012-2013 fishing year. Annual catches since 2012 are about 500 t .
The Group has in the past included the fraction of S. mentella that are caught with pelagic trawls above the western, southwestern and southern continental slope of Iceland as part of the landing statistics of the demersal S. mentella. This practice has been in accordance with Icelandic legislation, where captains are obligated to report their $S$. mentella catch as either "pelagic redfish" or as "demersal redfish" depending in which fishing area they fish. According to this legislation, all catch outside the Icelandic EEZ and west of the 'redfish line' (red line shown in Figure 18.1.1, which is drawn approximately over the 1000-m isoclines within the Icelandic EEZ) shall be reported as pelagic S. mentella. All fish caught east of the 'redfish line' shall be reported as demersal S. mentella. Most of the catches since 1991 have been taken by bottom trawlers along the shelf west, southwest, and southeast of Iceland at depths between 500 and 800 m . The Group accepts this praxis as a pragmatic management measure, but notes that there is no biological information that could support this catch allocation.

As the Review Group in 2005 noted that this issue needed more elaboration, detailed portrayals of the geographical, vertical and seasonal distribution of the demersal $S$. mentella fisheries with different gears are presented here, as done previously (see below). Quantitative information on the fractions of the pelagic catches of demersal $S$. mentella is given in chapter 20 . The proportion of the total demersal S. mentella catches taken by pelagic trawls has ranged since 1991 between $0 \%$ and $44 \%$ (Table 20.3.2), and is on average $15 \%$. With exception of 2007 , no demersal S. mentella has been caught with pelagic trawls since 2004. The geographic distribution of the Icelandic fishery for
S. mentella since 1991 was in general close to the redfish line, off South Iceland, and has expanded into the NAFO Convention Area since 2003 (Figure 18.4.1). The pelagic catches of demersal S. mentella were taken in similar areas and depths as the bottom trawl catches (Figure 18.4.2). The vertical and horizontal distribution of the pelagic catches focused, however, on smaller areas and shallower depth layers than the bot-tom-trawl catches. The seasonal distribution by depth (Figure 18.4.3) shows that the pelagic catches of demersal S. mentella were in general taken in autumn, and overlapped in June with the traditional pelagic fishery only in 2003 and 2007. The bottomtrawl catches of the demersal S. mentella were mainly taken in the first quarter of the year and during autumn/winter. The length distributions of the demersal S. mentella catches in Iceland by gear and area are given in Figure 18.4.4. During 1994-1999 and in 2003, the fish taken with pelagic trawls were considerably larger than the fish caught with bottom trawls, but they were of similar length during 2000-2002. The fish caught in the northeastern area were on average about 5 cm larger than those caught in the southwestern area. The length distribution also shows that the fish caught in this area since 2011 is smaller than during the period 1998-2010 and have now a size similar to that registered at the beginning of the fishery.

### 18.5 Russian pelagic S. mentel/a fishery

Russia's position regarding the structure of redfish stock in the Irminger Sea remains unchanged and it has been expressed in previous reports (ICES, 2009, Annex 4; ICES, 2013; Makhrov et al. 2011; Zelenina et al. 2011). The Russian Federation still maintains its point of view that there is only one stock of beaked redfish S. mentella in the pelagic waters of the Irminger Sea and that is why no split catches information about the fisheries is presented to the NWWG. Russia reiterates its standpoint that studies of the redfish stock structure should be continued (Artamonova et. al 2013) with the aim of developing agreed recommendations using all available scientific and fisheries data as a basis.

In 2015 the fishery was conducted from April to September in ICES Subareas 12 and 14 and NAFO Divisions 1F (Tables 21.2.1, 21.2.2, 22.2 .1 and 22.2.2) with average cpue 20.6 $\mathrm{t} /$ day and 20.7 t / day in ICES Subareas 12 and 14, respectively; and 36.7 t /day in NAFO.

### 18.6 Biological sampling

Biological samples are taken both in national and international surveys and from the commercial catches. They consist of length measurements, otolith collection, stomach contents, sex and maturity stages. The following samples were taken by several nations during 2015:

| COUNTRY | AREA | No. OF <br> SAMPLES | No. OF FISH MEASURED |
| :--- | :---: | :---: | :---: |
| Russia | 14 | 350 | 40,053 |
| Russia | 12 |  | 7,244 |
| Russia | NAFO 1F | 150 | 15,356 |
| Iceland | $14($ deep $)$ | 6 | 996 |
| Greenland | 14 b |  |  |

### 18.7 Demersal S. mentel/a in 5b and 6

### 18.7.1 Demersal S. mentel/a in 5b

### 18.7.1.1 Surveys

The Faroese spring and summer surveys in Division 5 b are mainly designed for species inhabiting depths down to 500 m and do not cover the vertical distribution of demersal S. mentella fully. Therefore, the surveys are not used to evaluate the stock status.

### 18.7.1.2 Fisheries

In Division 5b, landings gradually decreased from 15000 t in 1986 to about 5000 t in 2001 (Table 18.6.1). Between 2002 and 2011 annual landings varied between 1100 and 4,000 t. In 2012 landings decreased drastically and were about 500 t in 2015.

Length distributions from the landings in 2001-2015 indicate that the fish caught in 5b in 2015 are between $35-50 \mathrm{~cm}$ and the mode of the distribution is around 41 cm (Figure 18.7.1).

Non-standardized cpue indices in Division 5b were obtained from the Faroese otter board ( OB ) trawlers ( $>1000 \mathrm{HP}$ ) towing deeper than 450 m and where demersal S . mentella composed at least $70 \%$ of the total catch in each tow. The OB trawlers have in recent years landed about $50 \%$ of the total demersal S. mentella landings from Vb . cpue decreased from $500 \mathrm{~kg} / \mathrm{hour}$ in 1991 to $300 \mathrm{~kg} /$ hour in 1993 and remained at that level until 2013, when it reached a historical low (Figure 18.7.2). The cpue has since remained at that level.

Fishing effort has decreased since the beginning of the time-series and remains very low since 2008.

### 18.7.2 Demersal S. mente/la in 6

### 18.7.2.1 Fisheries

In Subarea 6, the annual landings varied between 200 t and 1100 t in 1978-2000 (Table 18.6.1). The landings from 6 in 2004 were negligible ( 6 t ), the lowest recorded since 1978. They increased again to 111 t in 2005 and 179 t in 2006. The reported landings in 2008 were 50 t and no catches have been taken since 2009.

### 18.8 Regulations (TAC, effort control, area closure, mesh size etc.)

Management of redfish differs between stock units and is described in sections 19.14 for $S$. norvegicus, section 20.7 for Icelandic slope S. mentella, section 21.10 for shallow pelagic S. mentella, section 22.10 for deep pelagic S. mentella, and section 23 for Greenland slope S. mentella

The allocation of Icelandic S. mentella catches to the pelagic and demersal management unit has been based on the "redfish line" (see section 18.4).

### 18.9 Mixed fisheries, capacity and effort

The official statistics reported to ICES do not divide catch by species/stocks, and since the Review Group in 2005 recommended that "multispecies catch tables are not relevant to management of redfish resources", these data are not given here and the best estimates on the landings by species/stock unit are given in the relevant chapters. Preliminary official landings data were provided by the ICES Secretariat, NEAFC and

NAFO, and various national data were reported to the Group. The Group, however, repeatedly faced problems in obtaining catch data, especially with respect to pelagic $S$. mentella (see chapter 19.11). Detailed descriptions of the fisheries are given in the respective chapters: S. norvegicus in chapter 19.3, demersal S. mentella in chapter 20.3, shallow pelagic S. mentella in chapter 21.2, deep pelagic S. mentella in chapter 22.2 and Greenland slope redfish in chapter 23.3.
Information from various sources is used to split demersal landings into two redfish species, S. norvegicus and S. mentella (see stock annexes for Icelandic slope S. mentella and S. norvegicus). In Division 5a, if no direct information is available on the catches for a given vessel, the landings are allocated based on logbooks and samples from the fishery. According to the proportion of biological samples from each cell (one fourth of ICES statistical square), the unknown catches within that cell are split accordingly and raised to the landings of a given vessel. For other areas, samples from the landings are used as basis for dividing the demersal redfish catches between $S$. norvegicus and $S$. mentella.

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### 18.11 Tables

Table 18.4.1. Landings of S. viviparus in Division 5a 1996-2014.

| YEAR | LANDINGS (T) |
| :---: | :---: |
| 1996 | 22 |
| 1997 | 1159 |
| 1998 | 994 |
| 1999 | 498 |
| 2000 | 227 |
| 2001 | 21 |
| 2002 | 20 |
| 2003 | 3 |
| 2004 | 2 |
| 2005 | 4 |
| 2006 | 9 |
| 2007 | 24 |
| 2008 | 15 |
| 2009 | 37 |
| 2010 | 2602 |
| 2011 | 1427 |
| 2012 | 535 |
| 2013 | 532 |
| 2014 | 550 |
| 2015 | 468 |

Table 18.6.1. Nominal landings (tonnes) of demersal S. mentella 1978-2015 in ICES Divisions5b and 6.

| Year |  |  |
| :---: | :---: | :---: |
| 1978 | 7767 | 18 |
| 1979 | 7869 | 819 |
| 1980 | 5119 | 1109 |
| 1981 | 4607 | 1008 |
| 1982 | 7631 | 626 |
| 1983 | 5990 | 396 |
| 1984 | 7704 | 609 |
| 1985 | 10560 | 247 |
| 1986 | 15176 | 242 |
| 1987 | 11395 | 478 |
| 1988 | 10488 | 590 |
| 1989 | 10928 | 424 |
| 1990 | 9330 | 348 |
| 1991 | 12897 | 273 |
| 1992 | 12533 | 134 |
| 1993 | 7801 | 346 |
| 1994 | 6899 | 642 |
| 1995 | 5670 | 536 |
| 1996 | 5337 | 1048 |
| 1997 | 4558 | 419 |
| 1998 | 4089 | 298 |
| 1999 | 5294 | 243 |
| 2000 | 4841 | 885 |
| 2001 | 4696 | 36 |
| 2002 | 2552 | 20 |
| 2003 | 2114 | 197 |
| 2004 | 3931 | 6 |
| 2005 | 1593 | 111 |
| 2006 | 3421 | 179 |
| 2007 | 1376 | 1 |
| 2008 | 750 | 50 |
| 2009 | 1077 | 0 |
| 2010 | 1202 | 0 |
| 2011 | 1126 | 0 |
| 2012 | 263 | 0 |
| 2013 | 398 | 0 |
| 2014 | 370 | 0 |
| $2015{ }^{1)}$ | 537 | 0 |

[^3]
### 18.12 Figures



Figure 18.1.1 Potential management unit boundaries. The polygon bounded by blue lines, i.e. 1, indicates the region for the 'deep pelagic' management unit in the northwest Irminger Sea, 2 is the "shallow pelagic" management unit in the southwest Irminger Sea, and 3 is the Icelandic slope management unit.


Figure 18.1.2 Schematic representation of biological stocks and potential management units of $S$. mentella in the Irminger Sea and adjacent waters. The management units are shown in Figure 18.1.1. Included is a schematic representation of the geographical catch distribution in recent years. Note that the shallow pelagic stock includes demersal S. mentella east of the Faroe Islands and the deep pelagic stock includes demersal S. mentella west of the Faroe Islands.


Figure 18.2.1 Survey abundance indices of Sebastes spp. $(<17 \mathrm{~cm})$ for East and West Greenland from the German groundfish survey 1982-2015.


Figure 18.4.1Geographical distribution of the Icelandic catches of S. mentella 1991-2002. The color scale indicates catches (tonnes per $\mathbf{N M}^{2}$ ).


Figure 18.4.1 cont. Geographical distribution of the Icelandic catches of S. mentella 2003-2015. The color scale indicates catches (tonnes per $\mathrm{NM}^{2}$ ),


Figure 18.4.2 Distance-depth plot for Icelandic S. mentella catches, where distance (in NM) from a fixed position $\left(52^{\circ} \mathrm{N} 50^{\circ} \mathrm{W}\right)$ is given. The contour lines indicate catches in a given area and distance. The colored contours represent the fishery on pelagic $S$. mentella, the black contours indicate bot-tom-trawl catches of demersal $S$. mentella, and the red contours represent catches of demersal $S$. mentella taken with pelagic trawls.


Figure 18.8.3 Depth-time plot for Icelandic S. mentella catches 1991-2015 where the y-axis is depth, the $x$-axis is day of the year and the color indicates the catches. The colored contours represent the fishery on pelagic S. mentella, the black contours indicate bottom-trawl catches of demersal S. mentella, and the red contours represent catches of demersal S. mentella taken with pelagic trawls.


Figure 18.8.4 Length distributions from different Icelandic S. mentella fisheries, 1991-2015. The blue lines represent the fishery on pelagic $S$. mentella in the northeastern area, the red lines the pelagic fishery in the southwestern area, the black lines indicate bottom-trawl catches of demersal S. mentella, and the green lines represent catches of demersal S. mentella taken with pelagic trawls.


Figure 18.7.1 Length distribution of demersal S. mentella from landings of the Faroese fleet in Division 5b 2001-2015.


Figure 18.7.2 Demersal S. mentella, cpue (t/hour) and fishing effort (in thousands hours) from the Faroese CUBA fleet 1991-2015 and where 70\% of the total catch was demersal S. mentella.

## 19 Golden redfish (Sebastes norvegicus) in Subareas 5, 6 and 14

### 19.1 Stock description and management units

Golden redfish (Sebastes norvegicus) in ICES Subareas 5 and 14 have been considered as one management unit.

Catches in ICES Subarea 6 have traditionally been included in this report and the Group continues to do so.

### 19.2 Scientific data

This chapter describes results from various surveys conducted annually on the continental shelves and slopes of Subareas 5 and 14

### 19.2.1 Division 5.a

Two bottom-trawl surveys are conducted in Icelandic waters: the Spring Survey in March 1985-2016 and the Autumn Survey in October 1996-2015. The autumn survey was not conducted in 2011. Two survey indices are calculated from these surveys and used in the assessment of golden redfish in ICES 5.a. Length disaggregated indices from the Spring Survey are used in the Gadget model. Age disaggregated indices from the autumn survey are used as age-length keys in 2 cm length groups in the Gadget model.

The survey stratification and subsequent survey indices for golden redfish were recalculated for the Autumn Survey in 2008 and for the Spring Survey in 2011. The method is described in the Stock Annex for the species. Further changes were made in the calculation of the survey indices in 2012 by taking into account length dependent diurnal vertical migration of the species. Golden redfish is known for its diurnal vertical migration showing semi-pelagic behaviour. Usually the species is in the pelagic area during the night-time and close to the bottom during the daytime. However, there is also a size or age difference in this pelagic behaviour where smaller fish shows opposite vertical migration pattern compared to larger fish. The method is described in more details in the Stock Annex.

This scaled diurnal variation by length was used for calculating Cochran index for redfish. The sum of those abundance indices multiplied by mean weight at length or age are the total indices shown in Figure 19.2.1 and Table 19.2.1.

Figure 19.2.1 $a$ shows the total biomass index from the Icelandic spring and autumn groundfish surveys with $\pm 1$ standard deviation in the estimate ( $68 \%$ confidence interval). The total biomass of golden redfish as observed in the spring survey decreased from 1988 to a record low in 1995. Between 1996 and 2002 the stock showed signs of improvement but was low compared to the beginning of the series. From 2003-2012 the biomass increased significantly, but decreased again in 2014 and 2015 although remained high. The total biomass index in 2016 increased substantially (about 50\% compared to 2015) and was the highest recorded. The index is twice as high as at the beginning of the time-series. The CV of the measurement error has been considerably higher since 2003 than before that.
The total biomass index from the autumn survey gradually increased from 2000-2014 when it was the highest in the time-series, but decreased slightly in 2015 (Figure 19.2.1a).

Length distribution from the spring survey shows that the peaks, which can be seen first in 1987 and then in 1991-1992, reached the fishable stock approximately 10 years later (Figure 19.2.2). The increase in the survey index between 1995 and 2005 reflects the recruitment of a relatively strong year classes (1985-year class and then the 1990year class). Abundance of small redfish has since then been much smaller, highest in 1998-2000, but in recent six years very little has been observed of small redfish (Figure 19.2.1d). This has been confirmed by age readings (Figure 19.2.4). In recent years the modes of the length distribution in both surveys has shifted to the right and is narrower. Much less is now observed of golden redfish less than 30 cm compared to other years (Figures 19.2.2 and 19.2.3).

Age disaggregated abundance indices from the autumn survey is shown in Figure 19.2.4 and Table 19.2.2. The sharp increase in the survey indices since 2005 reflects the recruitment of the year-classes from 1996-2005. The year-classes 1996-1999 are gradually disappearing from the stock. The 2000-2005 year-classes are now similar to the indices of the large 1990 year-class at same age. In 2013-2015, the abundance of fish 6 years' old and younger was at the lowest level in the time-series for all age groups (Table 19.2.2).

### 19.2.2 Division 5.b

In Division 5.b, cpue of S. norvegicus were available from the Faroes spring groundfish survey from 1994-2016 and the summer survey 1996-2015. Both surveys show similar trends in the indices from 1998 onwards with sharp declines between 1998 and 1999 (Figure 19.2.5). After an increase in the mid-1990s, cpue decreased drastically. cpue in the spring survey was between 2000 and 2008 stable at low level. In the period 20092016 it has been at the lowest level since the beginning of the series. The cpue index in the summer survey has gradually decreased and is also at the lowest level recorded.

### 19.2.3 Subarea 14

Relative abundance and biomass indices from the German groundfish survey from 1982-2015 for S. norvegicus (fish $>17 \mathrm{~cm}$ ) are illustrated in Figure 19.2.6. In 2013, the survey was re-stratified, with 4 strata in West Greenland resembling NAFO subarea structure, and 5 strata in East Greenland. Depth zones considered are 0-200 m and 200400 m . The time-series was recalculated accordingly. In general, the survey indices are much lower with the new stratification scheme but show similar trend (WD 30 of the 2013 NWWG report).
After a severe depletion of the $S$. norvegicus stock on the traditional fishing grounds around East Greenland in the early 1990's, the survey estimates showed a significant increase in both abundance and biomass with the highest value observed in 2007 (Figure 17.2.7). The survey indices were high although fluctuating until 2013. The survey index increased in 2014 to the highest level in the time-series and was almost two times higher than in 2013 (Figure 19.2.6a and Figure 19.2.6b). The index decreased in 2015 but was the second highest in the time-series. It should be noted that the CV for the indices are high and the increase is driven by few very large hauls. During the recent period of increase, both the fishable biomass ( $>30 \mathrm{~cm}$ ) and the biomass of pre-fishery recruits $(17-30 \mathrm{~cm})$ have increased considerably (Figures 19.2.7c and 19.2.8). In 2010-2015 the biomass of $17-30 \mathrm{~cm}$ fish has decreased compared to previous five years whereas the fishable biomass has remained high since 2007.

Abundance indices of redfish smaller than 18 cm from the German annual groundfish survey show that juveniles were abundant in 1993 and 1995-1998 (Figure 18.2.1). Since

2008, the survey index has been very low and was in 2015 the lowest value recorded since 1982. Juvenile redfish were only classified to the genus Sebastes spp., as species identification of small specimens is difficult due to very similar morphological features. The 1999-2015 survey results indicate low abundance and are similar to those observed in the late 1980s. The Greenland shrimp and fish shallow water survey also shows no juvenile redfish ( $<18 \mathrm{~cm}$, not classified to species) were present.

### 19.3 Information from the fishing industry

### 19.3.1 Landings

Total landings gradually decreased by more than $70 \%$ from about 130000 t in 1982 to about $43,000 \mathrm{t}$ in 1994 (Table 19.3.1 and Figure 19.3.1). Since then, the total annual landings have varied between 33,500 and 54,000 $t$. The total landings in 2015 were 51645 t , which is about 900 t more than in 2014. The majority of the golden redfish catch is taken in ICES Division 5.a that contributes to about $94-98 \%$ of the total landings.

Landings of golden redfish in Division 5.a declined from about 98000 t in 1982 to 39000 t in 1994 (Table 19.3.1). Since then, landings have varied between 32000 t and 51000 t , highest in 2013. The landings in 2015 were about 48800 t , about 1000 t more than in 2014. Between $90-95 \%$ of the golden redfish catch is taken by bottom trawlers targeting redfish (both fresh fish and factory trawlers; vessel length $48-65 \mathrm{~m}$ ). The remaining catches are partly caught as bycatch in gillnet, longline, and lobster fishery. In 2015, as in previous years, most of the catches were taken along the shelf southwest, west and northwest of Iceland (Figure 19.3.2). Larger proportion of the catches is now taken along the shelf northwest of Iceland and less south and southwest.
In Division 5.b, landings dropped gradually from 1985-1999 from $9000 \mathrm{t}-1500 \mathrm{t}$ and varied between 1500 and 2500 t from 1999-2005 (Table 19.3.1). In 2006-2015 annual landings were less than 1000 t which has not been observed before in the time-series. The landings in 2015 were 270 t which is 70 t more than in 2014. The landings in 2014 and 2015 are the lowest landings in the time-series. The majority of the golden redfish caught in Division 5b is taken by pair and single trawlers (vessels larger than 1000 HP ).
Annual landings from Subarea 14 have been more variable than in the other areas (Table 19.3.1). After the landings reached a record high of 31000 t in 1982, the golden redfish fishery drastically reduced within the next three years (the landings from ICES Subarea 14 were about 2000 t in 1985). During the period 1985-1994, the annual landings from Subarea 14 varied between 600 and $4,200 \mathrm{t}$, but from 1995 to 2009 there was little or no direct fishery for golden redfish and landings were 200 t or less mainly taken as bycatch in the shrimp fishery. In 2010, landings of golden redfish increased considerable and were 1650 t , similar to it was in early 1990s. This increase is mainly due to increased S. mentella fishery in the area. Annual landings 2010-2015 have been between 1000 t and 2700 t , highest in 2014.

Annual landings from Subarea 6 increased from 1978-1987 followed by a gradual decrease to 1992 (Table 19.3.1). From 1995 to 2004, annual landings have ranged between 400 and 800 t , but decreased to 137 t in 2005. Little or no landings of golden redfish were reported from Subarea 6 in 2006-2015 and were 44 t in 2015.

### 19.3.2 Discard

Comparison of sea and port samples from the Icelandic discard sampling program does not indicate significant discarding due to highgrading in recent years (Palsson et al 2010), possibly due to area closures of important nursery grounds west off Iceland.

Substantial discard of small redfish took place in the deep-water shrimp fishery from 1986-1992 when sorting grids became mandatory. Since then the discard has been insignificant both due to the sorting grid and much less abundance of small redfish in the region.

Discard of redfish species in the shrimp fishery in ICES Division 14.b is currently considered insignificant (see Chapter 18).

### 19.3.3 Biological data from the commercial fishery

The table below shows the fishery related sampling by gear type and ICES Divisions in 2015. No sampling of the commercial catch from subdivision VI was carried out.

| Area | Nation | Gear | LANDINGS (T) | Samples | No. <br> LENGTH MEASURED | No. Age <br> READ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5.a | Iceland | Bottom trawl | 48769 | 211 | 36756 | 1836 |
| 5.b | Faroe Islands | Bottom trawl | 270 | 8 | 248 |  |
| 14 | Greenland | Bottom trawl | 2457 |  |  |  |

### 19.3.4 Landings by length and age

The length distributions from the Icelandic commercial trawler fleet in 1976-2015 show that the majority of the fish caught is $30-45 \mathrm{~cm}$ (Figure 19.3.3). The modes of the length distributions range between 35 and 37 cm . The length distributions in 2012-2015 are narrower than previously, with less than average of both small and large fish caught.
Catch-at-age data from the Icelandic fishery in Division 5.a show that the 1985-year class dominated the catches from 1995-2002 (Figure 19.3.4 and Table 19.3.2) and in 2002 this year class still contributed to about $25 \%$ of the total catch in weight. The strong 1990-year class dominated the catch in 2003-2007 contributing between 25-30\% of the total catch in weight. The share of these two year classes has gradually been decreasing in recent years. In 2007-2010 the 1996-1999 year classes dominated in the catches, but are now gradually decreasing. The 2000-2004 year classes contributed in total about 62\% of the total catch in 2015.

The average total mortality $(\mathrm{Z})$, estimated from the 21-year series of catch-at-age data (Figure 19.3.5) is about 0.24 for age groups 15 and older.

Length distribution from the Faroese commercial catches for 2001-2015 indicates that the fish caught are on average larger than 40 cm with modes between 45 cm and 50 cm (Figure 19.3.6).

No length data from the catches have been available for several years in Subareas 14 and 6 .

### 19.3.5 cpue

The un-standardized cpue index was in 2015 the highest in the time-series with sharp increase in recent 8 years. Effort towards golden redfish has since 1986 gradually decreased and is at the lowest level recorded (Figure 19.3.7). cpue derived from logbooks is not considered indicative of stock trends however the information contained in the logbooks on effort, spatial and temporal distribution the fishery is of value.

Un-standardized cpue of the Faroese otter-board (OB) trawlers has been presented in previous reports. They are however considered unreliable and un-representative about
the stock in Division 5.b. This is because no separation of S. norvegicus/S. mentella is made in the catches.

### 19.4 Methods

### 19.4.1 Changes to the assessment model in January 2014.

The stock was benchmarked in January 2014 and a management plan evaluated and adopted (WKREDMP, ICES 2014). The benchmark group agreed to base the advice for next five years on the Gadget model. The settings are described in the Stock Annex. The following changes were done to the model compared to previous runs:

- Abundance indices from the German survey in East Greenland were included in the tuning. The indices were added to the Icelandic spring survey.
- Tuning data were limited to $19-54 \mathrm{~cm}$ instead of $25-54 \mathrm{~cm}$ as larger part of the stock area is included. 19 cm is around the length at which redfish in the German survey is classified to species. Earlier, smaller fish had gradually been removed from the tuning fleet as the nursery area for year classes 19962003 seemed to be outside Icelandic waters.
- Length at recruitment was estimated separately for year classes 1996-2000 and 2001 and onwards. The reason was higher mean weight at age in landings and autumn survey.

Of the changes mentioned above, the first one has the largest effect on the estimated stock size but the third one does also have considerable effect as when growth increases fish recruit to the fisheries at younger age if selection is size dependent.

The German survey did get half weight compared to the results in Figure 19.2.6. This was done to avoid extrapolation to areas not surveyed, and hence reduce noise, but the indices are calculated as numbers per square $\mathrm{km}^{2}$ multiplied by an area drawn around the stations (Figure 19.4.1). By using the stratification used to calculate indices shown in Figure 19.2.6, each station in the German survey would get 2.5 times more weight compared to the Icelandic survey. Several things are not comparable between the two surveys, for example different gears are used and the German survey is not conducted during night while the Icelandic survey is conducted both day and night. Therefore the "correct" weight of each survey in the total is difficult to estimate and part of the benchmark work 2014 was to look at the sensitivity to the weight.
The German survey has in recent decade provided increased proportion of the total biomass, but is still only $10 \%$ of the total biomass (Figure 19.4.2). The contribution for each length group (Figure 19.4.3) does though show that large redfish is abundant in East Greenland and large part of the largest redfish $(45+\mathrm{cm})$ is found there. This affects the model results as the relatively large abundance of middle size redfish in the Icelandic spring survey (Figure 19.2.1a) has not lead to subsequent increase in large fish (Figure 19.2.1c). Including the large fish from East Greenland does therefore affect model results and estimated SSB is $20 \%$ higher when the German survey is included, although the German survey does only account for $10 \%$ of the total biomass as it is weighted. The recruitment signal from the German survey (Figure 19.4.3) is on the other hand not explaining much of the "missing recruitment" from Icelandic waters in recent years.

The weighing of individual datasets in the Gadget model is done using an iterative reweighing algorithm. The process essentially assigns weights to each input dataset on the basis of the inverse variance of the fitted residuals. This is done to reduce the effect
of low quality input data. In this year assessment the weights were the same as in the benchmark runs in January 2014 and the assessment in 2014 and 2015.

### 19.4.2 Gadget model

### 19.4.2.1 Data and model settings

Below is a brief description of the data used in the model and model settings is given. A more detailed description is given in the Stock annex.

Data used in the Gadget model are:

- Length disaggregated survey indices $19-54 \mathrm{~cm}$ in 2 cm length increments from the Icelandic groundfish survey in March 1985-2016 and the German survey in East Greenland 1984-2015. Indices are added together and the German survey gets half the weight compared to what is presented in Figure 19.2.6.
- Length distributions from the Icelandic, Faroe Islands and East Greenland commercial catches since 1970.
- Landings by 6 month period from Iceland, Faroe Islands and East Greenland.
- Age-length keys and mean length-at-age from the Icelandic groundfish survey in October 1996-2015.
- Age-length keys and mean length-at-age from the Icelandic commercial catch 1995-2015.
- The simulation period is from 1970 to 2020 using data until the first half of 2016 for estimation. Two time-steps are used each year. The ages used were $5-30$ years, where the oldest age is treated as a plus group (fish 30 years and older). Recruitment was set at age 5 .

Estimated parameters are:

- Number of fish when the simulation starts (8 parameters).
- Recruitment-at-age 5 each year (44 parameters).
- Length at recruitment (3 parameters).
- Parameters in the growth equation; (2 parameters).
- Parameter $\beta$ of the beta-binomial distribution controlling the spread of the length distribution.
- Selection pattern of the three commercial fleets assuming logistic selection (S-shape) ( $3 \times 2$ parameters).
- Selection pattern of the survey fleet assuming an Andersen selection curve (bell-shape) (3 parameters).

It needs to be mentioned that the length disaggregated indices are from the spring survey but the age data are from the autumn survey conducted six months later. The surveys could have different catchability but the age data are used as proportions within each 2 cm length group so it should not matter. Growth in between March and October is taken care of by the model.

Projections were run using the Gadget model based fishing mortality of equal to 0.097 for ages 9-19 according to agreed management plan.
Assumptions done in the predictions:

- Recruitment-at-age 5 in 2015 and onwards was set as the average of the recruitment in 2012-2014.
- Catches in the first time-step in 2016 (first 6 months) were set at the same as in the first time-step of 2015 for all the fleets. In step 2 in 2016 and onwards the model was run at fixed effort corresponding to $\mathrm{F}_{9-1}=0.097$
- The estimated selection pattern from the Icelandic fleet was used for projections.


### 19.4.2.2 Results of the assessment model and predictions

Summary of the assessment is shown in Figure 19.4.4 and Table 19.4.1. The spawning stock has increased in recent years and fishing mortality decreased but annual landings have been relatively stable. The last year class estimated is the 2014 year class but the following year-classes are assumed to be the average of the 2012-2014 year classes. Compared to last year's assessment the 2007-2012 year-class is estimated larger than assumed last year. Later year-classes are likely to be smaller than assumed here based on information from the surveys in East Greenland and Iceland that all indicate low abundance of small redfish. Assumptions about those year-classes will not have much effect on the advice this year but later advice will be affected as well as the development of the spawning stock in short term.

The results of the assessment presented here are similar to what was presented at WKREDMP (ICES 2014) (Figure 19.4.5). This similarity is expected as only one year of data has been added and the model is a is a low pass filter that does usually not respond rapidly to new data except they are very far from predicted values.

Estimated selection patterns of different fleets are shown in Figure 19.4.6. The Greenlandic and Faroese fleet catch much larger fish than the Icelandic fleet. This is in line with the results from the German survey in East-Greenland that show most of the large fish in East-Greenland (Figure 19.4.3)

### 19.4.2.3 Fit to data

An aggregated fit to the survey index (converted to biomass) is presented in Figure 19.4.7. It shows a greater level of agreement than most runs based only on the Icelandic data but does mostly show negative residuals for the last 14 years. Residuals by length group show positive residuals in size groups $33-38 \mathrm{~cm}$ in recent years but negative for most other size groups, indicating narrower length distributions in the survey than predicted (Figure 19.5.8).

This lack of fit between observed and predicted survey biomass was one of the main critics of WKRED 2012 (ICES 2012). As can be seen in Figure 19.4.7 the fit is still not good. That lack of fit is caused by too narrow length distribution, with both small and large fish missing but they weight much more in the tuning data than in the total biomass. When looking at the number of years with observed > predicted biomass it must be noted that the assessment converges very slowly and 10 years are in some sense comparable to less than 5 years in other species. Discussions about the problem in WKRED 2014 are still valid.

The correlation between observed and predicted survey indices is good for $33-50 \mathrm{~cm}$ fish (Figures 19.4.9 and 19.4.10). As the model converges slowly, predicted indices could change a number of years back when more data are added. However, it is not the magnitude of the residuals but rather the temporal pattern that is worrying (Figure 19.4.8).

Length distributions from the Icelandic commercial catch does usually show good fit except in the most recent period when the large fish is missing and the length distribution narrower than ever (Figure 19.4.11). One explanation could be that selection in recent years is dome shaped as the large fish is in East Greenland where the fisheries are less.

The discrepancy between predicted and observed age distributions is not as apparent as for the length distributions (Figures 19.4.12 and 19.4.13). The model uses the data as age-length keys in 2 cm intervals for tuning. Presenting the residuals on that scale is difficult so here the age distributions are shown as aggregates overall length groups. This is not a problem for the catches where the otolith sampling is random, which is not the case for the survey as there is a maximum limit on the number of otoliths sampled in each tow and therefore smaller proportion sampled in hauls with many fish.

### 19.5 Information from catch curves.

The discrepancy in different data sources can be seen by looking at catch curves from age disaggregated catch in numbers and survey indices. The 1995-1999 year-classes have disappeared more rapidly from the fisheries than predicted with average $Z$ being $0.24(F=0.19)$ for ages $12-20$. Comparable number for year-classes $1985-1990$ is $Z=0.15$.

The analyses indicate that fishing mortality was higher than predicted by the assessment models. One explanation is that we are overestimating the stock but there can be a number of alternative explanations.

- The cohorts grow faster and mature earlier than earlier cohorts. Natural mortality, $M$, might have increased
- The selection of the fisheries is more dome shaped than before. The fisheries concentrate on the dense schools west of Iceland where the length distribution is narrow.
- Compared to cohorts 1985-1990 the later cohorts seem to come from other nursery areas.
- Most of the biomass in the Icelandic surveys in the last decade comes from very dense schools west of Iceland. Catchability in those schools might be different from less dense aggregations.


### 19.6 Reference points

Harvest control rule (HCR) was evaluated at WKREDMP in January 2014 (ICES, 2014) based on stochastic simulations using the Gadget model. Taking into account conflicting information by different data continuing for many consequent years (sections 19.419.5), the simulations were conducted using large assessment error with very high autocorrelation ( $\mathrm{CV}=0.25$, $\mathrm{rho}=0.9$ ).

Yield-per-recruit analysis show that when average size at age 5 was allowed to change after year class 1996, F9-19,max changed from 0.097 to 0.114 . The proposed fishing mortality of 0.097 is therefore around $85 \%$ of $\mathrm{F}_{\max }$ with current settings. Stochastic simulations indicate that it leads to very low probability of spawning stock going below $\mathrm{B}_{\text {trigger }}$ and Blim, even with relatively large autocorrelated assessment error.

The simulations done at WKREDMP 2014 (ICES, 2014) were repeated, but with deterministic recruitment and no assessment error. At WKREDMP 2014, Blim $=B_{l o s s}=160 \mathrm{kt}$ was defined as the lowest SSB in the 2012 Gadget run. $\mathrm{B}_{\text {trigger }}=\mathrm{B}_{\mathrm{pa}}$ was defined as 220 kt by
adding a precautionary buffer to the proposed $\operatorname{Blim}$ of 160 kt : $160^{*} \exp \left(0.2^{*} 1.645\right)$. Recruitment in the stochastic simulations was the average of year-classes 1975-2003 but those year-classes were the basis for the simulations at WKREDMP 2014.

The plot of the average spawning stock against fishing mortality show that $\mathrm{F}_{\mathrm{lim}}=0.226$ and $\mathrm{F}_{\mathrm{pa}}$ is then $0.226 / \exp \left(1.645^{*} 0.2\right)=0.163$ (Figure 19.6.1). The spawning stock decreased considerably from early 1980s to mid-1990s or from 400 kt to 200 kt . The reduction in SSB was due to heavy fisheries, but increased again gradually because of improved recruitment and lower F (Figure 19.6.1).

The probability of current $\mathrm{SSB}<\mathrm{B}_{\text {trigger }}$ is estimated $2.7 \%$. For simplicity, the action of Btrigger is not included in the simulations since Gadget is not keeping track of "perceived spawning stock". Analysis of the stochastic prediction in R shows that if SSB is below $B_{\text {trigger }}$ it will only be noted in $<15 \%$ of the cases. The reason is that the spawning stock is only likely to go below $B_{\text {trigger }}$ in periods of severe overestimation of the stock that occur due to the assumed high autocorrelation in assessment error. This situation differs from that of the stock going below $B_{\text {trigger }}$ due to poor recruitment (worse than observed in recent decades). In this case the spawning stock should still have a resilient age structure (as discussed above) and this could reduce the need to take further action below $B_{\text {trigger }}$.

Figure 19.6.2 shows the development of $\mathrm{F}_{9-19}$ based on $\mathrm{F}_{9-19}=0.097$. F is expected to be within the range of the fifth and 95th quantile and the 16th and 84th quantile.

### 19.7 State of the stock

The results from Gadget indicate that fishing mortality has reduced in recent years and is now close to $\mathrm{F}_{\text {msy }}$ (Figure 19.4.4). Spawning stock and fishable stock have been increasing in recent years and are now the highest since 1986.

In 5b, survey indices are stable at low level and do not indicate an improved situation in the area. In Subarea 14, the biomass of the fishable stock has been relatively high since 2007. No information is available on exploitation rates in Division 5.b and Subarea 14.

Results from surveys in Iceland and East Greenland indicate that most recent year classes are poor. The reliability of the surveys as an indicator of recruitment is not known.

### 19.8 Short-term forecast

The Gadget model is length based where growth is modelled based on estimated parameters. The only parameters needed for short-term forecast are assumptions about size of those cohorts that have not been seen in the surveys. These year classes were assumed to be the average of year classes 1975-2003 (Figure 19.4.4).

The results from the short-term simulations based on $\mathrm{F}_{9}-19$ is shown in Figure 19.4.4 and from short-term prognosis with varying fishing mortality in 2017 and 2018 in Table 19.4.2.

### 19.9 Medium term forecast

No medium term forecast was carried out.

### 19.10Uncertainties in assessment and forecast

Various factors regarding the uncertainty and modelling challenges are listed in the WKRED-2012 (ICES 2012) and WKREDMP-2014 (ICES 2014) reports. The main issues
relate to the lack of explanation of the Gadget model (or any model for that matter) to account for the increase of abundance in intermediate length groups in the Icelandic March survey. These factors were discussed in sections 19.4-19.6 but a short list is repeated below.

- Immigration of intermediate sized redfish in to 5.a, most likely from Greenland.
- Increased aggregation of redfish in areas closed to fishing. These areas on the western part of the Icelandic shelf make up most but not all of the increase in intermediate sized golden redfish in the Icelandic surveys. However eliminating the hauls from these areas in calculation of indices does to some extend reduce this increase.
- There are indications that growth of golden redfish has changed over time. This can be seen for example in the 2001 year class which is on average larger than fish of the same age in the earlier year classes (for example, the 19851990 year classes). Size at maturity has also decreased that could lead to growth ceasing earlier than before explaining lack of large fish in recent years


### 19.11 Comparison with previous assessment and forecast

The current assessment gives similar state of the stock compared to last year's assessment and the assessment presented at the benchmark 2014.

Management plans and evaluation, see chapter 19.6

### 19.12 Basis for advice

Harvest control rule accepted at WKREDMP 2014 (ICES 2014).

### 19.13 Management consideration

In 2009 a fishery targeting redfish was initiated in Subarea 14 with annual catches of between 7300 and 8500 t in 2010-2015. The fishery does not distinguish between species, but based on survey information, golden redfish is estimated to be between 1000 and 2700 , highest in 2014.
Redfish and cod in Subarea 14 are found in the same areas and depths and historically these species have been taken in the same fisheries. An increased redfish fishery may therefore affect cod. ICES currently advise that no fishery should take place on offshore cod in Greenland waters. ICES therefore recommend measures that will keep effort on cod low in the redfish fishery.

Greenland opened an offshore cod fishery in 2008. To protect spawning aggregations of cod present management measures in Greenland EEZ prohibits trawl fishery for cod north of $63^{\circ} \mathrm{N}$ latitude. Restrictions on cod bycatch in fisheries directed towards other demersal fish (i.e. redfish and Greenland halibut) provide some protection of cod, but additional measures such as a closure of potential redfish fisheries north of $63^{\circ} \mathrm{N}$ could be considered.

Subarea 14 is an important nursery area for the entire resource. Measures to protect juvenile in Subarea 14 should be continued (sorting grids in the shrimp fishery).

No formal agreement on the management of S. norvegicus exists among the three coastal states, Greenland, Iceland and the Faroe Islands. However, an agreement was made between Iceland and Greenland in October 2015 on the management of the
golden redfish fishery based on the management plan applied in 2014. The agreement is from 2016 to the end of 2018. The agreement states that each year $90 \%$ of the TAC is allocated to Iceland and $10 \%$ is allocated to Greenland. Furthermore, 350 t are allocated each year to other areas.
In Greenland and Iceland the fishery is regulated by a TAC and in the Faroe Islands by effort limitation. The regulation schemes of those states have previously resulted in catches well in excess of TACs advised by ICES.

### 19.14 Ecosystem consideration

Not evaluated for this stock.

### 19.15 Regulation and their effects

The separation of golden redfish and Icelandic slope $S$. mentella quota was implemented in the 2010/2011 fishing season.

In the late 1980's, Iceland introduced a sorting grid with a bar spacing of 22 mm in the shrimp fishery to reduce the bycatch of juveniles in the shrimp fishery north of Iceland. This was partly done to avoid redfish juveniles as a bycatch in the fishery, but also juveniles of other species. Since the large year classes of golden redfish disappeared out of the shrimp fishing area, there in the early 1990's, observers report small redfish as being negligible in the Icelandic shrimp fishery. If the sorting grids work where the abundance of redfish is high is a question but not a relevant problem at the moment in $5 . \mathrm{b}$ as abundance of small redfish is low and shrimp fisheries limited.

There is no minimum landing size of golden redfish in Division 5.a. However, if more than $20 \%$ of a catch observed on board is below 33 cm a small area can be closed temporarily. A large area west and southwest of Iceland is closed for fishing in order to protect young golden redfish.
There is no regulation of the golden redfish in Division 5.b.
Since 2002 it has been mandatory in the shrimp fishery in Subarea 14 to use sorting grids in order to reduce bycatches of juvenile redfish in the shrimp fishery.

### 19.16 Changes in fishing technology and fishing patterns

There have been no changes in the fishing technology and the fishing pattern of golden redfish in Subareas 5 and 14.

### 19.17Changes in the environment

No information available.

### 19.18References

ICES 2012. Report of the Benchmark Workshop on Redfish (WKRED 2012). ICES CM 2012/ACOM:48, 291 pp.
ICES 2014. Report of the Workshop on Redfish Management Plan Evaluation (WKREDMP). ICES CM 2014/ACOM:52, 269 pp .

Pálsson, Ó., Björnsson, H., Björnsson, E., Jóhannesson, G. and Ottesen P. 2010. Discards in demersal Icelandic fisheries 2009. Marine Research in Iceland 154.

### 19.19Tables

Table 19.2.1 Survey indices and CV of golden redfish from the spring survey 1985-2015 and the autumn survey 1996-2014.

|  | Spring Survey |  | Autumn Survey |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | BIomass | CV | Biomass | CV |
| 1985 | 308,148 | 0.095 |  |  |
| 1986 | 328,104 | 0.120 |  |  |
| 1987 | 322,281 | 0.122 |  |  |
| 1988 | 251,625 | 0.095 |  |  |
| 1989 | 281,253 | 0.122 |  |  |
| 1990 | 242,514 | 0.223 |  |  |
| 1991 | 199,251 | 0.114 |  |  |
| 1992 | 160,614 | 0.088 |  |  |
| 1993 | 179,378 | 0.130 |  |  |
| 1994 | 171,171 | 0.097 |  |  |
| 1995 | 146,127 | 0.102 |  |  |
| 1996 | 195,190 | 0.165 | 196,938 | 0.249 |
| 1997 | 212,254 | 0.216 | 118,785 | 0.282 |
| 1998 | 206,550 | 0.136 | 186,397 | 0.348 |
| 1999 | 297,014 | 0.143 | 261,868 | 0.311 |
| 2000 | 221,295 | 0.176 | 140,774 | 0.201 |
| 2001 | 192,749 | 0.176 | 176,532 | 0.156 |
| 2002 | 249,375 | 0.173 | 191,887 | 0.151 |
| 2003 | 334,011 | 0.161 | 199,374 | 0.159 |
| 2004 | 327,175 | 0.236 | 219,935 | 0.242 |
| 2005 | 310,708 | 0.129 | 229,105 | 0.240 |
| 2006 | 257,213 | 0.157 | 279,435 | 0.335 |
| 2007 | 339,975 | 0.224 | 220,157 | 0.251 |
| 2008 | 248,120 | 0.154 | 287,813 | 0.244 |
| 2009 | 302,501 | 0.253 | 293,618 | 0.283 |
| 2010 | 383,772 | 0.245 | 227,259 | 0.171 |
| 2011 | 401,870 | 0.235 |  |  |
| 2012 | 461,192 | 0.204 | 342,800 | 0.226 |
| 2013 | 457,737 | 0.177 | 310,332 | 0.157 |
| 2014 | 403,085 | 0.174 | 430,585 | 0.233 |
| 2015 | 407,051 | 0.280 | 360,878 | 0.175 |
| 2016 | 608,992 | 0.317 |  |  |

Table 19.2.2 Golden redfish in 5.a. Age disaggregated indices (in numbers) from the autumn groundfish survey 1996-2015. The survey was not conducted in 2011.

| $\begin{gathered} \text { Year/ } \\ \text { Age } \\ \hline \end{gathered}$ | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.3 | 1.0 | 3.7 | 3.3 | 0.8 | 0.4 | 0.1 | 0.0 | 0.0 | 0.1 | 0.2 | 0.1 | 0.0 | 0.1 | 0.0 |  | 0.0 | 0.0 | 0.0 | 0.0 |
| 2 | 2.4 | 0.2 | 1.5 | 3.3 | 1.7 | 1.0 | 1.0 | 0.6 | 0.2 | 0.1 | 0.6 | 1.3 | 0.3 | 0.3 | 0.0 |  | 0.0 | 0.0 | 0.3 | 0.1 |
| 3 | 0.7 | 2.2 | 0.9 | 3.3 | 1.4 | 2.0 | 1.5 | 1.1 | 1.0 | 0.2 | 0.7 | 1.2 | 2.5 | 0.4 | 1.7 |  | 0.1 | 0.0 | 0.3 | 0.6 |
| 4 | 1.6 | 1.6 | 2.3 | 1.5 | 1.6 | 2.4 | 6.1 | 1.1 | 1.8 | 1.0 | 0.5 | 1.1 | 2.7 | 4.6 | 0.3 |  | 1.1 | 0.2 | 0.1 | 0.2 |
| 5 | 8.4 | 2.2 | 0.9 | 4.7 | 1.2 | 5.4 | 5.7 | 12.1 | 3.2 | 4.2 | 5.0 | 2.1 | 4.1 | 12.2 | 4.3 |  | 3.9 | 1.1 | 0.8 | 0.1 |
| 6 | 40.4 | 6.9 | 3.5 | 2.8 | 7.7 | 2.1 | 11.6 | 17.4 | 28.1 | 4.8 | 6.8 | 10.2 | 7.7 | 11.6 | 14.3 |  | 3.1 | 4.0 | 1.7 | 1.2 |
| 7 | 11.4 | 22.4 | 16.7 | 10.5 | 6.6 | 10.6 | 3.2 | 37.5 | 35.9 | 39.0 | 15.2 | 25.5 | 38.3 | 13.7 | 15.0 |  | 23.0 | 3.0 | 12.6 | 7.3 |
| 8 | 19.1 | 14.2 | 58.4 | 47.2 | 6.2 | 10.6 | 26.3 | 9.7 | 63.8 | 43.9 | 79.7 | 35.0 | 73.1 | 72.3 | 23.0 |  | 68.6 | 40.7 | 23.9 | 27.4 |
| 9 | 15.0 | 12.9 | 22.5 | 100.0 | 25.5 | 6.9 | 10.9 | 47.5 | 20.4 | 61.2 | 79.2 | 74.7 | 65.8 | 94.1 | 53.4 |  | 58.9 | 82.2 | 93.7 | 32.0 |
| 10 | 28.9 | 11.0 | 26.1 | 43.7 | 92.7 | 16.9 | 16.1 | 12.4 | 44.3 | 24.3 | 83.2 | 36.5 | 103.3 | 57.1 | 67.8 |  | 61.1 | 54.1 | 146.8 | 83.5 |
| 11 | 102.6 | 17.4 | 18.9 | 20.6 | 11.1 | 108.5 | 31.1 | 16.6 | 18.7 | 43.1 | 25.6 | 35.2 | 61.3 | 98.1 | 31.9 |  | 100.8 | 39.4 | 87.7 | 97.5 |
| 12 | 15.9 | 67.0 | 19.0 | 16.7 | 13.8 | 22.9 | 113.5 | 39.0 | 13.0 | 19.1 | 36.4 | 18.6 | 53.5 | 44.6 | 56.5 |  | 72.0 | 65.2 | 67.2 | 51.1 |
| 13 | 9.9 | 6.0 | 104.8 | 20.7 | 7.7 | 23.0 | 19.5 | 109.3 | 25.9 | 15.0 | 17.5 | 23.2 | 13.2 | 41.7 | 28.2 |  | 42.1 | 45.2 | 65.3 | 46.0 |
| 14 | 16.6 | 5.2 | 10.0 | 147.2 | 7.8 | 7.6 | 11.2 | 12.1 | 101.2 | 26.2 | 14.7 | 8.0 | 17.8 | 9.9 | 19.3 |  | 38.0 | 25.1 | 48.8 | 40.3 |
| 15 | 33.6 | 7.0 | 7.6 | 5.9 | 50.4 | 8.9 | 9.6 | 10.7 | 13.3 | 80.5 | 17.9 | 6.6 | 8.9 | 17.8 | 9.0 |  | 19.1 | 30.1 | 26.1 | 39.1 |
| 16 | 15.9 | 9.8 | 7.8 | 9.6 | 5.1 | 57.4 | 10.2 | 6.0 | 9.4 | 9.3 | 73.7 | 16.6 | 7.6 | 6.7 | 10.8 |  | 16.2 | 17.8 | 25.7 | 20.4 |
| 17 | 1.8 | 6.8 | 14.1 | 10.8 | 2.5 | 4.1 | 44.3 | 7.5 | 5.8 | 6.5 | 8.5 | 48.6 | 12.8 | 6.2 | 4.6 |  | 6.0 | 12.2 | 16.5 | 19.4 |
| 18 | 1.6 | 3.8 | 7.6 | 11.0 | 2.5 | 4.9 | 4.5 | 32.3 | 5.9 | 3.7 | 4.2 | 10.2 | 35.9 | 7.1 | 3.0 |  | 5.7 | 6.8 | 11.9 | 9.7 |
| 19 | 4.1 | 2.0 | 0.5 | 8.3 | 4.5 | 3.5 | 2.9 | 4.5 | 21.1 | 5.0 | 2.7 | 4.4 | 6.0 | 27.6 | 6.5 |  | 3.8 | 4.9 | 5.8 | 9.7 |
| 20 | 6.5 | 1.4 | 3.2 | 3.9 | 6.4 | 4.0 | 3.1 | 1.6 | 3.0 | 21.7 | 3.0 | 1.5 | 5.6 | 4.5 | 21.9 |  | 3.8 | 4.3 | 5.7 | 9.6 |
| 21 | 1.0 | 0.8 | 2.3 | 2.8 | 1.0 | 3.6 | 3.9 | 1.1 | 1.8 | 2.5 | 17.5 | 3.9 | 2.0 | 2.1 | 3.1 |  | 3.4 | 4.6 | 4.7 | 3.2 |
| $22$ | 4.9 | $1.5$ | 0.8 | 1.0 | 1.6 | 2.2 | 3.2 | 2.7 | 1.7 | 2.0 | 1.9 | 13.5 | 2.3 | 1.3 | 1.2 |  | 17.9 | 2.3 | 3.5 | 2.4 |
| 23 | 3.9 | 2.4 | 2.2 | 2.1 | 0.4 | 0.3 | 0.8 | 1.0 | 2.4 | 2.3 | 1.7 | 1.3 | 10.8 | 1.9 | 1.6 |  | 2.9 | 17.3 | 3.3 | 2.1 |
| 24 | 4.5 | 0.8 | 0.4 | 0.5 | 1.0 | 0.5 | 0.4 | 0.3 | 0.0 | 0.9 | 1.0 | 1.2 | 1.4 | 9.9 | 0.7 |  | 2.0 | 2.4 | 12.3 | 1.1 |
| $25$ | 3.8 | $2.6$ | $1.4$ | 2.8 | 0.7 | $0.3$ | 0.5 | 0.3 | 1.2 | 1.2 | 1.7 | 0.2 | 0.8 | 0.7 | 5.6 |  | 1.2 | 1.2 | 1.4 | 12.7 |
| 26 | 0.8 | 1.1 | 0.2 | 1.1 | 0.6 | 0.5 | 0.5 | 0.2 | 0.4 | 0.3 | 0.9 | 0.6 | 0.8 | 0.9 | 0.6 |  | 1.6 | 1.1 | 0.9 | 1.5 |
| 27 | 0.8 | 0.2 | 0.9 | 2.9 | 0.5 | 0.7 | 0.3 | 0.3 | 0.0 | 0.1 | 0.9 | 0.3 | 1.2 | 1.2 | 0.4 |  | 7.4 | 0.8 | 0.8 | 1.4 |
| 28 | 0.8 | 0.4 | 0.5 | 1.5 | 0.6 | 0.5 | 0.2 | 0.0 | 0.2 | 0.2 | 0.2 | 0.0 | 0.5 | 0.2 | 0.7 |  | 0.4 | 8.3 | 0.5 | 1.5 |
| $29$ | 0.1 | $0.0$ | $0.4$ | $1.2$ | $0.5$ | $0.2$ | 0.7 | $0.1$ | 0.2 | 0.0 | 0.4 | 0.4 | 0.8 | 1.5 | 0.4 |  | 0.4 | 0.4 | 3.2 | 1.0 |
| 30+ | 0.8 | 1.3 | 3.1 | 1.1 | 1.3 | 2.1 | 1.6 | 1.5 | 1.5 | 2.1 | 1.0 | 0.9 | 1.4 | 1.6 | 2.0 |  | 2.0 | 3.3 | 2.5 | 6.7 |
| Total | 358.1 | 212.1 | 342.2 | 492.0 | 265.4 | 314.0 | 344.5 | 386.5 | 425.4 | 420.5 | 502.5 | 382.9 | 542.4 | 551.9 | 387.8 |  | 566.5 | 478.0 | 674.0 | 528.8 |

Table 19.3.1 Official landings (in tonnes) of golden redfish, by area, 1978-2015 as officially reported to ICES. Landings statistics for 2015 are provisional.

| Area |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 5.A | 5.B | 6 | 14 | Total |
| 1978 | 31300 | 2039 | 313 | 15477 | 49129 |
| 1979 | 56616 | 4805 | 6 | 15787 | 77214 |
| 1980 | 62052 | 4920 | 2 | 22203 | 89177 |
| 1981 | 75828 | 2538 | 3 | 23608 | 101977 |
| 1982 | 97899 | 1810 | 28 | 30692 | 130429 |
| 1983 | 87412 | 3394 | 60 | 15636 | 106502 |
| 1984 | 84766 | 6228 | 86 | 5040 | 96120 |
| 1985 | 67312 | 9194 | 245 | 2117 | 78868 |
| 1986 | 67772 | 6300 | 288 | 2988 | 77348 |
| 1987 | 69212 | 6143 | 576 | 1196 | 77127 |
| 1988 | 80472 | 5020 | 533 | 3964 | 89989 |
| 1989 | 51852 | 4140 | 373 | 685 | 57050 |
| 1990 | 63156 | 2407 | 382 | 687 | 66632 |
| 1991 | 49677 | 2140 | 292 | 4255 | 56364 |
| 1992 | 51464 | 3460 | 40 | 746 | 55710 |
| 1993 | 45890 | 2621 | 101 | 1738 | 50350 |
| 1994 | 38669 | 2274 | 129 | 1443 | 42515 |
| 1995 | 41516 | 2581 | 606 | 62 | 44765 |
| 1996 | 33558 | 2316 | 664 | 59 | 36597 |
| 1997 | 36342 | 2839 | 542 | 37 | 39761 |
| 1998 | 36771 | 2565 | 379 | 109 | 39825 |
| 1999 | 39824 | 1436 | 773 | 7 | 42040 |
| 2000 | 41187 | 1498 | 776 | 89 | 43550 |
| 2001 | 35067 | 1631 | 535 | 93 | 37326 |
| 2002 | 48570 | 1941 | 392 | 189 | 51092 |
| 2003 | 36577 | 1459 | 968 | 215 | 39220 |
| 2004 | 31686 | 1139 | 519 | 107 | 33451 |
| 2005 | 42593 | 2484 | 137 | 115 | 45329 |
| 2006 | 41521 | 656 | 0 | 34 | 42211 |
| 2007 | 38364 | 689 | 0 | 83 | 39134 |
| 2008 | 45538 | 569 | 64 | 80 | 46251 |
| 2009 | 38442 | 462 | 50 | 224 | 39177 |
| 2010 | 36155 | 620 | 220 | 1653 | 38648 |
| 2011 | 43773 | 493 | 83 | 1005 | 45354 |
| 2012 | 43089 | 491 | 41 | 2017 | 45635 |
| 2013 | 51330 | 372 | 92 | 1499 | 53263 |
| 2014 | 47769 | 201 | 60 | 2706 | 50736 |
| 2015 ${ }^{1)}$ | 48769 | 270 | 44 | 2562 | 51645 |

[^4]Table 19.3.2 Golden redfish in 5.a. Observed catch in weight (tonnes) by age and years in 1995-2015. It should be noted that the catch-at-age results for 1996 are only based on three samples, which explains that there are no specimens older than 23 years.

| Year/ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| 7 | 47 | 0 | 32 | 23 | 6 | 38 | 117 | 125 | 189 | 216 | 219 | 175 | 128 | 211 | 106 | 59 | 140 | 71 | 31 | 229 | 0 |
| 8 | 327 | 354 | 219 | 277 | 339 | 62 | 134 | 871 | 199 | 822 | 737 | 995 | 428 | 1,051 | 961 | 351 | 550 | 627 | 572 | 465 | 539 |
| 9 | 1,452 | 803 | 470 | 584 | 1,576 | 830 | 389 | 737 | 1,330 | 485 | 1,840 | 2,113 | 1,689 | 2,101 | 1,730 | 2,179 | 1,545 | 1,642 | 2,256 | 1,715 | 804 |
| 10 | 8,698 | 3,654 | 1,014 | 1,189 | 1,237 | 4,216 | 1,608 | 815 | 1,095 | 2,059 | 1,470 | 3,573 | 2,403 | 5,012 | 3,119 | 2,685 | 4,492 | 3,504 | 3,954 | 5,931 | 3,428 |
| 11 | 2,583 | 9,026 | 2,641 | 1,115 | 1,823 | 1,861 | 7,611 | 3,097 | 1,178 | 777 | 3,052 | 2,077 | 3,273 | 3,990 | 5,030 | 2,751 | 5,435 | 6,808 | 6,008 | 6,543 | 7,105 |
| 12 | 1,284 | 2,078 | 11,406 | 3,215 | 2,498 | 2,245 | 1,786 | 10,777 | 3,899 | 965 | 1,873 | 2,774 | 1,886 | 4,710 | 4,482 | 4,875 | 4,866 | 7,324 | 9,423 | 5,748 | 7,522 |
| 13 | 3,574 | 1,313 | 2,796 | 12,421 | 2,428 | 1,678 | 1,912 | 3,021 | 9,675 | 2,001 | 1,349 | 1,622 | 3,039 | 2,309 | 3,421 | 3,865 | 6,248 | 4,014 | 6,897 | 5,806 | 5,624 |
| 14 | 5,718 | 1,468 | 1,363 | 2,073 | 15,444 | 2,344 | 1,235 | 2,571 | 2,342 | 8,548 | 2,984 | 1,287 | 1,042 | 2,820 | 1,829 | 2,724 | 3,815 | 4,582 | 4,087 | 4,725 | 5,661 |
| 15 | 6,124 | 4,376 | 3,125 | 2,031 | 1,236 | 14,675 | 826 | 1,823 | 1,960 | 2,127 | 11,727 | 2,813 | 949 | 1,519 | 1,981 | 1,373 | 2,464 | 2,606 | 4,494 | 2,990 | 4,533 |
| 16 | 1,801 | 5,533 | 3,648 | 2,408 | 1,254 | 1,753 | 11,529 | 2,956 | 1,212 | 1,677 | 2,067 | 10,126 | 2,155 | 1,082 | 1,233 | 1,194 | 1,383 | 1,527 | 3,080 | 2,608 | 2,738 |
| 17 | 889 | 927 | 3,016 | 3,407 | 1,812 | 1,172 | 518 | 11,787 | 2,249 | 809 | 1,445 | 2,091 | 9,323 | 1,843 | 667 | 814 | 916 | 830 | 1,747 | 1,946 | 2,618 |
| 18 | 384 | 385 | 893 | 2,043 | 2,641 | 1,592 | 780 | 2,055 | 6,402 | 1,380 | 1,249 | 1,182 | 1,323 | 8,265 | 1,488 | 645 | 640 | 797 | 1,218 | 1,282 | 1,757 |
| 19 | 1,218 | 266 | 637 | 1,015 | 2,212 | 2,383 | 1,043 | 1,133 | 756 | 5,194 | 1,246 | 688 | 741 | 1,515 | 6,064 | 1,084 | 808 | 494 | 776 | 410 | 682 |
| 20 | 1,216 | 339 | 943 | 723 | 1,259 | 2,124 | 1,730 | 636 | 411 | 1,115 | 6,463 | 970 | 726 | 925 | 947 | 5,002 | 846 | 789 | 459 | 1,214 | 1,256 |
| 21 | 559 | 1,188 | 453 | 520 | 461 | 535 | 935 | 1,392 | 607 | 336 | 391 | 5,641 | 878 | 531 | 641 | 906 | 5,174 | 612 | 523 | 525 | 274 |
| 22 | 684 | 1,034 | 525 | 394 | 214 | 438 | 411 | 1,003 | 798 | 489 | 469 | 631 | 4,809 | 837 | 568 | 762 | 1,173 | 3,460 | 714 | 531 | 274 |
| 23 | 1,574 | 814 | 673 | 424 | 331 | 270 | 411 | 723 | 754 | 618 | 795 | 229 | 736 | 4,235 | 335 | 574 | 761 | 456 | 3,176 | 538 | 233 |
| 24 | 709 | 0 | 584 | 660 | 216 | 63 | 164 | 372 | 392 | 567 | 619 | 377 | 112 | 380 | 2,529 | 667 | 221 | 340 | 190 | 3,204 | 475 |
| 25 | 824 | 0 | 734 | 520 | 848 | 392 | 123 | 288 | 300 | 258 | 420 | 472 | 618 | 253 | 97 | 2165 | 67 | 226 | 201 | 201 | 1,845 |
| 26 | 407 | 0 | 275 | 399 | 270 | 337 | 114 | 180 | 74 | 105 | 100 | 73 | 333 | 427 | 96 | 267 | 1,602 | 238 | 173 | 209 | 276 |
| 27 | 384 | 0 | 139 | 427 | 615 | 198 | 275 | 80 | 83 | 183 | 279 | 263 | 349 | 340 | 191 | 389 | 86 | 1,441 | 74 | 116 | 179 |
| 28 | 808 | 0 | 202 | 357 | 229 | 516 | 189 | 296 | 27 | 141 | 169 | 204 | 200 | 170 | 92 | 132 | 178 | 200 | 822 | 64 | 188 |
| 29 | 0 | 0 | 143 | 53 | 106 | 364 | 146 | 498 | 105 | 138 | 29 | 168 | 36 | 172 | 386 | 179 | 47 | 73 | 38 | 733 | 78 |
| 30+ | 251 | 0 | 408 | 493 | 768 | 1,102 | 1,080 | 1,333 | 539 | 678 | 1,599 | 976 | 1,187 | 841 | 448 | 511 | 317 | 427 | 417 | 35 | 681 |

Table 19.4.1 Results from the Gadget model of total biomass, spawning-stock biomass, re-cruitment-at-age 5, catch and fishing mortality, projections are in italic.

| Year | Biomass | SSB | R(AGE5) | Catches | F9-19 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 607.6 | 373.0 | 214.4 | 67.9 | 0.098 |
| 1972 | 607.0 | 367.5 | 159.3 | 50.9 | 0.075 |
| 1973 | 649.8 | 375.6 | 457.9 | 43.7 | 0.065 |
| 1974 | 683.5 | 389.1 | 222.1 | 50.6 | 0.073 |
| 1975 | 703.1 | 398.7 | 135.4 | 61.9 | 0.087 |
| 1976 | 708.1 | 395.9 | 201.6 | 94.4 | 0.133 |
| 1977 | 717.2 | 399.6 | 185.3 | 53.8 | 0.079 |
| 1978 | 744.2 | 423.3 | 125.5 | 48.7 | 0.065 |
| 1979 | 761.8 | 440.3 | 163.6 | 77.2 | 0.099 |
| 1980 | 751.9 | 441.9 | 105.3 | 89.1 | 0.113 |
| 1981 | 722.4 | 432.1 | 75.0 | 102.0 | 0.134 |
| 1982 | 665.3 | 403.1 | 63.6 | 130.3 | 0.183 |
| 1983 | 599.8 | 366.6 | 67.7 | 106.0 | 0.161 |
| 1984 | 547.0 | 337.6 | 73.3 | 95.3 | 0.154 |
| 1985 | 509.8 | 314.3 | 131.5 | 78.5 | 0.131 |
| 1986 | 479.7 | 294.6 | 121.7 | 76.9 | 0.140 |
| 1987 | 443.7 | 272.3 | 64.5 | 76.6 | 0.151 |
| 1988 | 396.1 | 241.7 | 41.3 | 89.8 | 0.204 |
| 1989 | 355.4 | 215.4 | 45.0 | 56.6 | 0.144 |
| 1990 | 355.4 | 199.5 | 353.3 | 66.3 | 0.191 |
| 1991 | 334.2 | 182.5 | 59.0 | 56.0 | 0.179 |
| 1992 | 315.5 | 168.9 | 40.1 | 55.8 | 0.196 |
| 1993 | 299.2 | 157.8 | 54.7 | 50.2 | 0.194 |
| 1994 | 289.0 | 151.9 | 64.9 | 42.5 | 0.172 |
| 1995 | 308.3 | 151.6 | 337.9 | 44.3 | 0.181 |
| 1996 | 314.3 | 154.1 | 90.0 | 35.6 | 0.143 |
| 1997 | 314.3 | 156.2 | 41.7 | 39.0 | 0.153 |
| 1998 | 316.3 | 161.2 | 42.0 | 39.7 | 0.153 |
| 1999 | 314.3 | 162.5 | 86.9 | 42.5 | 0.162 |
| 2000 | 309.9 | 164.5 | 53.3 | 42.6 | 0.158 |
| 2001 | 317.2 | 169.0 | 116.4 | 36.7 | 0.130 |
| 2002 | 321.2 | 170.2 | 126.7 | 50.7 | 0.177 |
| 2003 | 338.0 | 174.5 | 194.7 | 38.2 | 0.133 |
| 2004 | 356.6 | 186.2 | 115.0 | 32.8 | 0.110 |
| 2005 | 379.5 | 195.6 | 187.0 | 46.6 | 0.153 |
| 2006 | 407.5 | 207.0 | 197.5 | 42.1 | 0.138 |
| 2007 | 425.6 | 219.9 | 113.9 | 39.2 | 0.123 |
| 2008 | 453.2 | 238.7 | 142.2 | 46.2 | 0.136 |
| 2009 | 490.0 | 257.8 | 231.6 | 39.3 | 0.108 |
| 2010 | 528.9 | 285.3 | 161.0 | 38.5 | 0.096 |
| 2011 | 548.3 | 309.9 | 52.1 | 45.1 | 0.103 |
| 2012 | 559.2 | 326.6 | 91.6 | 45.2 | 0.096 |
| 2013 | 562.7 | 342.7 | 44.9 | 53.1 | 0.106 |
| 2014 | 550.3 | 348.7 | 28.0 | 50.8 | 0.097 |
| 2015 | 541.9 | 354.6 | 54.0 | 51.8 | 0.096 |
| 2016 | 527.9 | 354.8 | 54.0 | 53.0 | 0.098 |
| 2017 | 509.1 | 349.8 | 54.0 | 52.8 | 0.099 |
| 2018 | 488.3 | 341.2 | 54.0 | 51.5 | 0.099 |
| 2019 | 466.9 | 330.1 | 54.0 | 49.6 | 0.099 |
| 2020 | 445.5 | 317.3 | 54.0 | 47.3 | 0.099 |

Table 19.4.2 Output from short-term prognosis. Multiplier is based on reference to the adopted HCR F $_{9-19}=0.097$. Biomasses are at the beginning of the year to apply to ICES standard in short-term prognosis in other places in the report they are in the middle of the year.

| $F(2015)=0.101 \mathrm{C}(2015)=51.000$ tons. |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 |  |  |  |  |  |  |
| BıO $5+$ |  | SSB | Fmult | F9-19 |  | Landings |
| 532 |  | 371 | 0.997 | 0.097 |  | 53.0 |
|  |  | 2017 |  |  | $2018$ |  |
| Fmult | F9-19 | BıO 5+ | SSB | Landings | Bıo 5+ | SSB |
| 0.0 | 0 | 542 | 383 | 0 | 576 | 389 |
| 0.1 | 0.01 | 539 | 383 | 5.7 | 568 | 389 |
| 0.2 | 0.019 | 536 | 383 | 11.4 | 559 | 389 |
| 0.3 | 0.029 | 534 | 383 | 16.9 | 551 | 389 |
| 0.4 | 0.038 | 531 | 383 | 22.3 | 542 | 389 |
| 0.5 | 0.048 | 528 | 383 | 27.6 | 534 | 389 |
| 0.6 | 0.058 | 525 | 383 | 32.9 | 526 | 389 |
| 0.7 | $0.067$ | 523 | 383 | 38.0 | 518 | 389 |
| 0.8 | 0.077 | 520 | 383 | 43.0 | 510 | 389 |
| 0.9 | 0.087 | 517 | 383 | 48.0 | 502 | 389 |
| 1.0 | 0.097 | 515 | 383 | 52.8 | 495 | 389 |
| 1.1 | 0.107 | 512 | 383 | 57.6 | 487 | 389 |
| 1.2 | 0.117 | 509 | 383 | 62.3 | 480 | 389 |
| 1.3 | 0.127 | 507 | 383 | 66.8 | 472 | 389 |
| 1.4 | 0.137 | 504 | 383 | 71.3 | 465 | 389 |
| 1.5 | 0.147 | 501 | 383 | 75.7 | 457 | 389 |
| 1.6 | 0.158 | 498 | 383 | 80.0 | 450 | 389 |
| 1.7 | 0.168 | 496 | 383 | 84.2 | 443 | 389 |
| 1.8 | 0.178 | 493 | 383 | 88.3 | 436 | 389 |
| 1.9 | 0.189 | 490 | 383 | 92.4 | 429 | 389 |
| 2.0 | 0.199 | 488 | 383 | 96.3 | 423 | 389 |

### 19.20Figures



Figure 19.2.1 Indices of golden redfish in ICES Division 5.a (Icelandic waters) from the groundfish surveys in March 1985-2016 (blue line and shaded area) and October 1996-2015 (red lines and shaded areas). The shaded areas show $\pm 1$ standard error of the estimate.


Figure 19.2.2. Length disaggregated abundance indices of golden redfish from the bottom trawl survey in March 1985-2016 conducted in Icelandic waters. The black line is the mean of total indices 1985-2015.


Figure 19.2.3. Length disaggregated abundance indices of golden redfish from the bottom-trawl survey in October 1996-2015 conducted in Icelandic waters. The black line is the mean of total indices 1996-2014. The survey was not conducted in 2011.


Figure 19.2.4 Age disaggregated abundance indices of golden redfish in the bottom-trawl survey in October conducted in Icelandic waters 1996-2015. The survey was not conducted in 2011.


Figure 19.2.5 cpue of golden redfish in the Faroes spring groundfish survey 1994-2016 and the summer groundfish survey 1996-2015 in ICES Division 5.b.


Figure 19.2.6 Golden redfish ( $>17 \mathrm{~cm}$ ). Survey abundance indices for East Greenland (ICES Subarea 14) from the German groundfish survey 1985-2015. a) Total biomass index, b) total abundance index, c) biomass index divided by size classes (17-30 cm and $>30 \mathrm{~cm}$ ).


Figure 19.2.7 Golden redfish ( $>17 \mathrm{~cm}$ ). Length frequencies for East Greenland (ICES Subarea 14) 1982-2015.


Figure 19.3.1 Nominal landings of golden redfish in tonnes by ICES Divisions 1978-2015. Landings statistics for 2015 are provisional.


Figure 19.3.2 Geographical distribution of golden redfish bottom-trawl catches in Division 5.a 2002-2015.


Figure 19.3.3 Length distribution (grey shaded area) of golden redfish in Icelandic waters (ICES Division 5.a) in the commercial landings of the Icelandic bottom-trawl fleet 1976-2015. The blue line is the mean of the years 1976-2014.


Figure 19.3.4 Catch-at-age of golden redfish in numbers in ICES Subdivision 5.a 1995-2015.


Figure 19.3.5 Catch curve of golden redfish based on the catch-at-age data in ICES Division 5.a 19952015.


Figure 19.3.6 Length distribution of golden redfish from Faroese catches in ICES Division 5.b in 2001-2015.


Figure 19.3.7 cpue of golden redfish from Icelandic trawlers 1978-2015 where golden redfish catch composed at least $50 \%$ of the total catch in each haul (black line), $80 \%$ of the total catch (red line) and in all tows where golden redfish was caught (blue line). The figure shows the raw cpue index (sum(yield)/sum(effort)) and effort.


Figure 19.4.1 Stations in the German survey in East Greenland with an area used to compile the indices for Gadget shown. This area corresponds to giving a weight of 0.5 to the results in Figure 19.2.7.


Figure 19.4.2 Biomass index from Iceland (blue) and Greenland black, based on weighting the German survey data in Figure 19.2.7 by 0.5.


Figure 19.4.3. Indices from the Icelandic March survey (red) and Icelandic March survey plus German survey in Greenland (blue) by length group.


Figure 19.4.4. Summary from the assessment. Red values are predictions. Spawning stock is compiled using a fixed maturity ogive with $L 50=33 \mathrm{~cm}$.


Figure 19.4.5. Comparison of the current assessment and the same assessment done in 2014 and 2015.


Figure 19.4.6. Estimates of selection curves from commercial catches (upper panel) and from the Icelandic March survey. The black line is the estimated selection curve fitted to the length distributional data and the red line is the estimated $q$ from the disaggregated tuning indices, scaled to one.


Figure 19.4.7. Comparison of observed and predicted survey biomass from the 2014 (red line) and 2015 (blue) runs.


Figure 19.4.8. Residuals from the fit between model and survey indices. The red circles indicate positive residuals (survey results exceed model prediction). Largest residuals correspond to $\log ($ obs $/$ mod $)=1$


Figure 19.4.9. Fit to length disaggregated survey indices from Gadget run as XY-scatter. The red line is fitted going through the 0 -point, the green cross goes over the terminal year.


Figure 19.4.10. Fit (red lines) to length disaggregated survey indices (broken lines and points) from Gadget run as time-series.


Figure 19.4.11. Fit (red line) to Icelandic commercial length distributions aggregated by 3 years.


Figure 19.4.12. Fit to survey age data (run 1). Bars represent the data and red lines the fit. The likelihood data are used in the model as proportions in each 2 cm length group but presented here as total for each age group something that should only be comparable if catchability was independent of size (age).


Figure 19.4.13. Predicted (red) and observed (blue) age distributions from Icelandic commercial fishery.


Figure 19.6.1. Average SSB against average fishing mortality and defined reference points.


Figure 19.6.2. Development of $\mathrm{F}_{9-19}$ based on $\mathrm{F}_{9-19}=0.097$. The light grey area shows fifth and 95th quantile and the dark areas 16th and 84th quantile.

## 20 Icelandic slope Sebastes mentella in 5.a and 14

### 20.1 Stock description and management units

The stock structure of Sebastes mentella in the Irminger Sea and adjacent water is described in Chapter 18 and Stock Annex. The S. mentella on the continental shelf and slope of Iceland is treated as separate biological stock and management unit. Only the fishable stock of Icelandic slope S. mentella is found in Icelandic waters, i.e. mainly fish larger than 30 cm . The East Greenland shelf is most likely a common nursery area for the three biological stocks described in Stock Annex, including the Icelandic slope one.

### 20.2 Scientific data

Only the fishable stock of Icelandic slope S. mentella is found in Icelandic waters. The Icelandic autumn survey on the continental shelf and slope in Division 5.a, covering depths down to 1500 m , does, therefore, not cover the whole distribution of the stock. Data for Icelandic slope S. mentella from the Autumn Survey is available from 20002015. No survey was conducted in 2011. A description of the autumn survey is given in Stock Annex for the species.

The total biomass index and the abundance indices from the autumn survey were highest in 2001. After a decrease in 2003 the index increased again in 2006 but gradually decreased until 2013 and to similar level as in 2003 when it was lowest in the timeseries (Table 20.2.1 and Figure 20.2.1 $a$ and $b$ ). The biomass index increased again and was in 2014 and 2015 similar as in 2004 (Table 20.2.1 and Figure 20.2.1a and b). The biomass index of fish 45 cm and larger was at lowest level in 2007 but increased again in 2009 were it was at similar level until 2013 (Figure 19.2.1c). The biomass index of 45 cm and larger fish increased sharply since then and was in 2015 the highest in the timeseries.

The abundance index of fish 30 cm and smaller has in 2007-2015 been at lowest level (Figure 20.2.2d). The length of the Icelandic slope $S$. mentella in the autumn survey is between 25 and more than 50 cm . Since 2000, the mode has shifted to the right, that is, from $36-39 \mathrm{~cm}$ in 2000 to about $42-43 \mathrm{~cm}$ in 2012-2015 (Figure 20.2.2). Very little Icelandic slope $S$. mentella smaller than 35 cm was observed in the surveys in recent years.
Otoliths have been sampled since 2000 and otoliths from the 2000, 2009 and 2010 surveys have been age read. Figure 20.2.3 shows that the 1985 and the 1990 year classes are the most abundant ones in this samples.

### 20.3 Information from the fishing industry

### 20.3.1 Landings

Total annual landings of Icelandic slope S. mentella from ICES Division 5.a 1978-2015 are presented in Table 20.3.1 and from 1950-2015 in Figure 20.3.1. Annual landings gradually decreased from a record high of 57000 t in 1994 to 17000 t in 2001. Landings in 2001-2010 fluctuated between 17000 t and 20500 except in 2003 and 2008 when annual landings were 28500 t and 24000 respectively. The landings in 2013-2015 between $8700-9500 \mathrm{t}$ and the decrease is related to lower TAC for the species.

### 20.3.2 Fisheries and fleets

Most of the fishery for Icelandic slope S. mentella in 5.a is a directed bottom-trawl fishery taken by bottom trawlers along the shelf and slope west, southwest, and southeast
of Iceland at depths between 500 and 800 m (Figure 20.3.2). The proportion of Icelandic slope S. mentella catches taken by pelagic trawls 1991-2000 varied between 10 and $44 \%$ of the total landings (Table 20.3.2). In 2001-2015, no pelagic fishery occurred or it was negligible except in 2003 and 2007 (see Stock Annex). In general, the pelagic fishery was mainly in the same areas as the bottom-trawl fishery (Figure 20.3.3).

A notable change in the catch pattern is that catches taken in the southeast fishing area has been gradually decreasing since 2000 and in recent years very little Icelandic slope S. mentella was taken on these fishing grounds (Figure 20.3.2). This area has historically been an important fishing area for Icelandic slope S. mentella.

### 20.3.3 Sampling from the commercial fishery

The table below shows the 2015 biological sampling from the catch and landings of Icelandic slope S. mentella in ICES Division Va. This is considered to be adequate sampling from the fishery. Otoliths from the commercial catch have been collected, but no systematic age reading is done.

| YeAR | NATION | Gear | LANDINGS (T) | No. SAMPLES | No. LENGTH <br> MEASURED |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $5 . a$ | Iceland | Bottom trawl | 9311 | 53 | 8530 |

### 20.3.4 Length distribution from the commercial catch

Length distributions of Icelandic slope S. mentella in 5.a from the bottom-trawl fishery show an increase in the number of small fish in the catch in 1994 compared to previous years (Figure 20.3.5). The peak of about 32 cm in 1994 can be followed by approximately 1 cm annual growth in 1996-2002. The fish caught in 2004-2014 peaked around $39-42 \mathrm{~cm}$. The length distribution of Icelandic slope $S$. mentella from the pelagic fishery, where available, showed that in most years the fish was on average bigger than taken in the bottom-trawl fishery (Figure 20.3.5).

### 20.3.5 Catch per unit of effort

Trends in raw cpue and effort are shown in Figure 20.3.6. cpue gradually decreased from 1978 to a record low in 1994, but has since then slightly increased annually to 2000. The cpue estimate in 2015 was at similar level as in late 1980s and about $40 \%$ higher than it was in 1994. The cpue has been stable since 2010. From 1991-1994, when cpue decreased, the fishing effort increased drastically. Since then, effort decreased and is now at similar level as in the early 1980s.

### 20.3.6 Discard

Although no direct measurements are available on discards, it is believed that there are no significant discards of Icelandic slope S. mentella in the Icelandic redfish fishery.

### 20.4 Methods

No analytical assessment was conducted on this stock.

### 20.5 Reference points

There are no biological reference points for the species. Previous reference points established were based upon commercial cpue indices, but are now considered to be unreliable indicators of stock size. ICES has withdrawn these reference points.

Icelandic slope beaked redfish in ICES Division 5.a has previously been assessed based on trends in survey biomass indices from the Icelandic Autumn survey or in ICES "trends based assessment". Supplementary data used in the assessment includes information from the fishery and length distributions from the commercial catch and the Autumn Survey. ICES advised in 2013, based on DLS approach (Method 3.2), that catches are set no higher than 9875 t in 2014. Same advice was applied for the 2015 fishing year, but was 10775 t for 2016. Annual TAC set by the Icelandic government was 10000 t in 2014-2016.

### 20.6 State of the stock

The Group concludes that the state of the stock is on a low level. With the information at hand, current exploitation rates cannot be evaluated for the Icelandic slope S. mentella in Division 5.a.

The fishable biomass index of Icelandic slope $S$. mentella from the Icelandic autumn survey shows that the biomass index for 2004-2013 has decreased to similar level as in 2003 when it was at lowest level, but increased again in 2014 and 2015. The survey was not conducted in 2011. Standardized cpue indices show a reduction from highs in the late 1980s, but there is an indication that the stock has started a slow recovery since the middle of 1990s, when cpue was close to $50 \%$ of the maximum. The cpue index gradually increased from 1995-2010 to a similar level as in the late 1980s and has since then been at that level.

In 2000-2008, good recruitment was been observed in the German survey on the East Greenland shelf (growth of about $2 \mathrm{~cm} / \mathrm{yr}$ ) which is assumed to contribute to both the Icelandic slope and pelagic stock at unknown shares. The German survey and the Greenland shrimp and fish shallow water survey both show no new recruits ( $>18 \mathrm{~cm}$ ) and no juveniles are present ( $<18 \mathrm{~cm}$ ). This suggests that the fishery in coming years will be based on the same cohorts.

### 20.7 Management considerations

S. mentella is a slow growing, late maturing deep-sea species and is therefore considered vulnerable to overexploitation and advice has to be conservative.
The cpue has slightly increased annually since a record low in 1994, especially in recent $3-4$ years and is now $40 \%$ higher than in 1994. It is, however, not known to what extent cpue series reflect change in stock status of Icelandic slope S. mentella. The nature of the redfish fishery is targeting schools of fish using advancing technology. The effect of technological advances is to increase cpue, but is unlikely to reflect biomass increase.

The advice for 2008-2012 was that a management plan to be developed and implemented which takes into account the uncertainties in science and the properties of the fisheries. ICES suggested that catches of S. mentella are set no higher than 10000 t as a starting point for the adaptive part of the management plan. The advice for 2014 and 2015 were 9875 t based on the DLS approach (Category 3.2).
The Icelandic slope $S$. mentella fishery southeast of Iceland has gradually ceased since 2000 and very little fishing is conducted in this area. This fishing area was prior to 2000 very important fishing area for Icelandic slope S. mentella.

The landings increased in Division 5.a between 2002-2003 by about 10000 t when the fishery of pelagic S. mentella merged with the Icelandic slope fishery at the redfish line. Those two fisheries merged again in 2007.

There are no explicit management for Icelandic slope $S$. mentella but the species is within the TAC system described in Chapter 7.5. Icelandic authorities gave until the 2010/2011 a joint quota for golden redfish and Icelandic slope S. mentella in Icelandic waters, but now give separate quotas for the species.

### 20.8 Basis for advice

Icelandic slope $S$. mentella is considered a data limited stock (DLS) and should follow the ICES framework for such (Category 3.2). Below is the description of the formulation of the advice for the 2017 fishing year.

Based on the North Western Working Group recommendation, the stock is treated as a stock with survey data, but no proxies for MSY Btrigger or F values, are known. This means that the catch advice for 2017 is based on the survey adjusted status quo catch equation:

$$
C_{y+1}=C_{y-1}\left(\frac{\sum_{i=y-x}^{y-1} I_{i} / x}{\sum_{i=y-z}^{y-x-1} I_{i} /(z-x)}\right)
$$

where $I$ is the survey index, $x$ is the number of years in the survey average, $\mathrm{z}=5$ and $\mathrm{C}_{\mathrm{y}}$ 1 is the advice last year. The biomass is estimated to have increased by $41.2 \%$ between average of 2011-2013 (no survey conducted in 2011, but was set to the average of 2010 and 2012) and 2014 and 2015 (average of the two years). This implies an increase of catches of $41.2 \%$ in relation to the last year advise (10 775 t ), corresponding to catch of no more than 15211 t . A precautionary buffer of $20 \%$ consistent with the ICES approach is applied which gives catch of no more than 12922 t .

### 20.9 Regulation and their effects

There are no explicit management for Icelandic slope $S$. mentella. The species is managed under the ITQ system (see Chapter 7.5.1). Icelandic authorities gave until the 2010/2011 fishing year a joint quota for golden redfish (S. norvegicus) and Icelandic slope $S$. mentella. The separation of quotas was implemented in the fishing year that started September 1, 2010.

A general description of management and regulation of fish populations in Icelandic waters is given in Chapter 7.5 and in Stock Annex A. 2 with emphasis on Icelandic slope S. mentella where applicable.

### 20.10Tables

Table 20.2.1 Total biomass index of Icelandic slope S. mentella in the Icelandic Autumn Groundfish survey 2000-2015. No survey was conducted in 2011.

| YEAR | ICELAND | CV |
| :--- | :---: | :---: |
| 2000 | 138924 | 0.145 |
| 2001 | 164030 | 0.172 |
| 2002 | 96923 | 0.137 |
| 2003 | 64621 | 0.127 |
| 2004 | 98373 | 0.164 |
| 2005 | 114953 | 0.249 |
| 2006 | 124509 | 0.172 |
| 2007 | 85469 | 0.183 |
| 2008 | 82703 | 0.139 |
| 2009 | 99767 | 0.183 |
| 2010 | 81963 | 0.149 |
| 2011 |  |  |
| 2012 | 78016 | 0.144 |
| 2013 | 70250 | 0.139 |
| 2014 | 104307 | 0.185 |
| 2015 | 110512 | 0.176 |

Table 20.3.1 Nominal landings (in tonnes) of Icelandic slope S. mentella 1978-2015 ICES Division 5.a.

| Year | ICELAND | Others | Total |
| :---: | :---: | :---: | :---: |
| 1978 | 3693 | 209 | 3902 |
| 1979 | 7448 | 246 | 7694 |
| 1980 | 9849 | 348 | 10197 |
| 1981 | 19242 | 447 | 19689 |
| 1982 | 18279 | 213 | 18492 |
| 1983 | 36585 | 530 | 37115 |
| 1984 | 24271 | 222 | 24493 |
| 1985 | 24580 | 188 | 24768 |
| 1986 | 18750 | 148 | 18898 |
| 1987 | 19132 | 161 | 19293 |
| 1988 | 14177 | 113 | 14290 |
| 1989 | 40013 | 256 | 40269 |
| 1990 | 28214 | 215 | 28429 |
| 1991 | 47378 | 273 | 47651 |
| 1992 | 43414 | 0 | 43414 |
| 1993 | 51221 | 0 | 51221 |
| 1994 | 56674 | 46 | 56720 |
| 1995 | 48479 | 229 | 48708 |
| 1996 | 34508 | 233 | 34741 |
| 1997 | 37876 | 0 | 37876 |
| 1998 | 32841 | 284 | 33125 |
| 1999 | 27475 | 1115 | 28590 |
| 2000 | 30185 | 1208 | 31393 |
| 2001 | 15415 | 1815 | 17230 |
| 2002 | 17870 | 1175 | 19045 |
| 2003 | 26295 | 2183 | 28478 |
| 2004 | 16226 | 1338 | 17564 |
| 2005 | 19109 | 1454 | 20563 |
| 2006 | 16339 | 869 | 17208 |
| 2007 | 17091 | 282 | 17373 |
| 2008 | 24123 | 0 | 24123 |
| 2009 | 19430 | 0 | 19430 |
| 2010 | 17642 | 0 | 17642 |
| 2011 | 11738 | 0 | 11738 |
| 2012 | 11965 | 0 | 11965 |
| 2013 | 8761 | 0 | 8761 |
| 2014 | 9500 | 0 | 9500 |
| 20151) | 9311 | 0 | 9311 |

1) Provisional

Table 20.3.2 Proportion of the landings of Icelandic slope S. mentella taken in ICES Division 5.a by pelagic and bottom trawls 1991-2015

| YeAR | PELAGIC TRAWL | BOtTOM TRAWL |
| :---: | :---: | :---: |
| 1991 | $22 \%$ | $78 \%$ |
| 1992 | $27 \%$ | $73 \%$ |
| 1993 | $32 \%$ | $68 \%$ |
| 1994 | $44 \%$ | $56 \%$ |
| 1995 | $36 \%$ | $64 \%$ |
| 1996 | $31 \%$ | $69 \%$ |
| 1997 | $11 \%$ | $89 \%$ |
| 1998 | $37 \%$ | $63 \%$ |
| 1999 | $10 \%$ | $90 \%$ |
| 2000 | $24 \%$ | $76 \%$ |
| 2001 | $3 \%$ | $97 \%$ |
| 2002 | $3 \%$ | $97 \%$ |
| 2003 | $28 \%$ | $72 \%$ |
| 2004 | $0 \%$ | $100 \%$ |
| 2005 | $0 \%$ | $100 \%$ |
| 2006 | $0 \%$ | $100 \%$ |
| 2007 | $17 \%$ | $83 \%$ |
| 2008 | $0 \%$ | $100 \%$ |
| 2009 | $0 \%$ | $100 \%$ |
| 2010 | $0 \%$ | $100 \%$ |
| 2011 | $0 \%$ | $100 \%$ |
| 2012 | $0 \%$ | $100 \%$ |
| 2013 | $0 \%$ | $100 \%$ |
| 2014 | $0 \%$ | $100 \%$ |
| 2015 | $0 \%$ | $100 \%$ |
|  |  |  |
|  |  |  |

### 20.1 1 Figures



Figure 20.2.1 Survey indices of the Icelandic slope $S$. mentella in the autumn survey in ICES Division 5.a 2000-2015. No survey was conducted in 2011. The figure shows the total biomass index, total abundance index in millions of fish, biomass index of fish 45 cm and larger and abundance index of fish 30 cm and smaller.


Figure 20.2.2 Length distribution of Icelandic slope S. mentella in the Autumn Groundfish Survey in October 2000-2015 in ICES Division 5.a. No survey was conducted in 2011. The blue line is the mean of 2000-2014.


Figure 20.2.3 Age distribution of Icelandic slope S. mentella from the Autumn Survey in 2000 ( $\mathrm{n}=$ $1405), 2009(n=1101)$, and $2010(n=1206)$. The age class 50 are the combined age classes of 50 years and older.


Figure 20.3.1 Nominal landings (in tonnes) of Icelandic slope S. mentella from Icelandic waters (ICES Division 5.a and Subarea 14) 1950-2015.


Figure 20.3.2 Geographical location of the Icelandic slope S. mentella catches in Icelandic waters (ICES Division 5.a and Subarea 14) 1993-2015 as reported in logbooks of the Icelandic fleet using bottom trawl. The blue line indicates part of the management unit for the deep-pelagic redfish stock. The dotted line represents the 500 m isobaths.


Figure 20.3.3 Geographical location of the Icelandic slope S. mentella catches in Icelandic waters (ICES Division 5.a and Subarea 14) 1991-2003 and 2007 as reported in logbooks of the Icelandic fleet using pelagic trawl. The blue line indicates part of the proposed management unit for the deep-pelagic redfish stock. The dotted line represents the 500 m isobaths.


Figure 20.3.5 Length distributions of Icelandic slope S. mentella from the Icelandic landings taken with bottom trawl (blue line) and pelagic trawl (red line) in Icelandic waters (ICES Division 5.a and Subarea 14) 1991-2015.


Figure 20.3 .6 cpue and effort of Icelandic slope S. mentella from the Icelandic bottom-trawl fishery in Icelandic waters (ICES Division 5.a and Subarea 14) 1978-2015.


Figure 20.5.1. Icelandic slope S. mentella. Number aged plotted on log-scale. Grey lines correspond to $Z=0.1$ and $Z=0.3$.

## 21 Shallow Pelagic Sebastes mentella

### 21.1 Stock description and management unit

This section addresses the fishery for shallow pelagic S. mentella in the Irminger Sea and adjacent areas (parts of Division 5.a, Subareas 12 and 14; eastern parts of NAFO Divisions 1F, 2H and 2J) at depths shallower than 500 m .
The following text table summarizes the available information from fishing fleets in the Irminger Sea and adjacent waters fishing for the shallow pelagic redfish in 2015. Only Russia conducted directed fishery on the stock. It should be noted that they also fished the deep pelagic stock:

| RUSSIA |
| :--- |

### 21.2 Summary of the development of the fishery

The historic development of the fishery can be found in the Stock Annex. The clear changes in the spatial pattern of the fishery can be seen in Figure 21.2.1, based on logbook data from the Faroe Islands, Greenland, Iceland and Norway. A summary of the catches by ICES Divisions/NAFO regulatory area as estimated by the Working Group is given in Table 21.2.1 and Figure 21.2.2. The estimated catch for 2015 is 5595 t , a decrease from the 6423 t caught in 2014. The catches were almost entirely produced by Russia with 5434 t from ICES 12 and NAFO 1F (Tables 21.2.1 and 21.2.2).

There are no new cpue data for 2015. The standardized cpue index trend for the period 1994-2006 is shown in Figure 21.2.3. This standardized cpue series includes data from Faroe Islands, Iceland, Germany, Greenland, and Norway, and it is estimated with a GLM model including the factors year, ship, month and towing time. The model output is shown in Table 21.2.3 and the residuals are in Figure 21.2.4.

### 21.3 Biological information

There are no new data. The length distributions for the period 1989-2006 of biological stocks based on Icelandic data are shown in Figure 21.3.1. The length of the largest proportion of caught fish oscillates around 35 cm for the whole period.

### 21.4 Discards

Redfish form aggregations composed of individuals with a narrow size range, which results in very clean catches. Thus, discards are negligible according to available data from various institutes.

### 21.5 Illegal Unregulated and Unreported Fishing (IUU)

The Group had again difficulties in obtaining catch estimates from several fleets. Furthermore, there are problems with misreported catches from some nations. The Group requests NEAFC and NAFO to provide ICES in time with all the necessary information.

### 21.6 Surveys

The last international trawl-acoustic survey for the shallow pelagic stock was carried out in 2013 and it is described in detail in ICES WGRS Report 2013 (ICES, 2013). The next survey was scheduled to be carried out in June/July 2015 (ICES, 2013) but after Russia withdrew its participation it was not possible to cover the whole distribution of the stocks. Therefore, no new biomass estimates are available.

### 21.6.1 Survey acoustic data

Since 1994, the results of the acoustic survey show a drastic decreasing trend from 2.2 million $t$ to 600000 t in 1999 and have fluctuated between $700000 \mathrm{t}-90000 \mathrm{t}$ in 20012013 (Table 21.6.1). The 2003 estimate, however, was considered to be inconsistent with the time-series due to a shift in the timing of the survey.
The most recent trawl-acoustic survey on pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters was carried out by Iceland, Germany and Russia in June/July 2013. Approximately $341000 \mathrm{NM}^{2}$ were covered. Figures 20.6 .1 and 20.6 .2 show the biomass estimates for depth shallower than the DSL (Depth Scattering Layer). A total biomass of 91000 t was estimated acoustically in the layer shallower than the DSL (Table 21.6.1 and Figure 21.6.4). The results showed a substantial biomass decline in subarea B compared to 2011 but in other areas the biomass was similar as in 2011 (Table 21.6.2 and Figure 21.6.5 for area definition). Biological samples from the acoustic estimate within the DSL and shallower than 500 m showed a mean length of 36.0 cm (Figure 21.6.6).

### 21.6.2 Survey trawl estimates

In addition to the acoustic measurements, redfish biomass was estimated by correlating catches and acoustic values at depths shallower than 500 m at 200000 t , a $45 \%$ decrease respect the estimation of 360,000 for 2011 (Table 21.6.1 and Figure 21.6.4). Figure 21.6 .3 shows the distribution of the redfish catches within the DSL and shallower than 500 m . It should be noted that the estimate for 2013 was recalculated due to technical error made in 2013 (ICES 2014).

The obtained correlation was used to convert the trawl data at greater depths to acoustic values and from there to abundance. For that purpose, standardized trawl hauls were carried out at depth 350-500 m, evenly distributed over the survey area (Figure 21.6.3). For the time being, the correlation between the catch and acoustic values is based on few data points only and it is highly variable. It is also assumed that the catchability of the trawl is the same, regardless of the trawling depth, thus the abundance estimate obtained is questionable and must only be considered as a rough attempt to measure the abundance within the DSL. Evaluation on the consistency of the method has to wait until more data points are available.
Biological samples from the trawls taken at depth $<500 \mathrm{~m}$ showed a mean length of 35.5 cm . Figure 21.6 .3 shows the spatial distribution of samples used in the survey and Figure 21.6.6 shows the corresponding length distribution.

### 21.6.3 Methods

The assessment of pelagic redfish in the Irminger Sea and adjacent waters is based on survey indices, catches, cpue and biological data. See Stock Annex and Section 21.6 for details.

### 21.6.4 Reference points

For pelagic redfish in the Irminger Sea and adjacent waters, no analytical assessment is carried out due to data uncertainties and the lack of reliable age data. Thus, no reference points can be derived.

### 21.7 State of the stock

### 21.7.1 Short-term forecast

For pelagic redfish in the Irminger Sea and adjacent waters, no analytical assessment is carried out due to data uncertainties and the lack of reliable age data. Thus, no shortterm forecasts can be derived.

### 21.7.2 Uncertainties in assessment and forecast

### 21.7.2.1 Data considerations

Preliminary official landings data were provided by the ICES Secretariat, NEAFC and NAFO, and various national data were reported to the Group. The Group, however, repeatedly faces problems to obtain reliable catch data due to unreported catches of pelagic redfish and lack of catch data disaggregated by depth from some countries. There are indications that reported effort (and consequently landings) could represent only around $80 \%$ of the real effort in certain years (see Chapter 20.3.3 in the 2008 NWWG report, ICES, 2008). No new data in IUU have been available since 2008.
As in previous years, detailed descriptions on the horizontal, vertical and seasonal distribution of the fisheries were given.
The need and importance of having catch and biological data disaggregated by depth from all nations taking part in the fishery cannot be stressed strongly enough, and the Group urges all nations involved on supplying better data. With this need in mind, ICES sent a data call to all EU countries participating in the redfish fishery, encouraging stockholders to deliver detailed catch data before the WG would meet, but the response was very limited.

### 21.7.2.2 Assessment quality

The results of the international trawl-acoustic survey are given in section 21.6. Given the high variability of the correlation between trawl and acoustic estimates as well as the assumptions that need to be made about constant catchability across depth and areas, the uncertainty of these estimates is very high.
The reduction in biomass observed in the surveys within the hydroacoustic layer (about 2 million $t$ in the last decade) cannot be explained by the reported removal by the fisheries (about 500,000 $t$ in the entire depth range in 1995-2013) alone. A decreasing trend in the relative biomass indices in the acoustic layer, however, is visible since 1991.

It is not known to what extent cpue reflects changes in the stock status of pelagic $S$. mentella, since the fishery focuses on aggregations. Therefore, stable or increasing cpue series might not indicate or reflect actual trends in stock size, although decreasing cpue indices are likely to reflect a decreasing stock. The new data available to the NWWG were insufficient to estimate the cpue for 2013.

NEAFC set for 2015 a 0 TAC for Shallow Pelagic S. mentella. However, the Russian Federation decided on a unilateral quota of 27300 t . This quota was taken from both Shallow and Deep pelagic stocks, since the Russian Federation does not agree on the division of the $S$. mentella management units and stock structure.

### 21.7.3 Comparison with previous assessment and forecast

The data available for evaluating the stock status are similar to last year.

### 21.7.4 Management considerations

The Group needs more and better data and requests that NEAFC and NAFO provide ICES with all information leading to more reliable catch statistics.

The main feature of the fishery since 1998 is a clear distinction between two widely separated fishing grounds with pelagic redfish fished at different seasons and different depths. Since 2000, the southwestern fishing grounds extended also into the NAFO Convention Area. Biological data, however, suggest that the aggregations in the NAFO Convention Area do not constitute a separate stock. The NAFO Scientific Council agreed with this conclusion (NAFO, 2005). The Group concludes that at this time there are not enough scientific bases available to propose an appropriate split of the total TAC among the two fisheries/areas.

### 21.7.5 Ecosystem considerations

The fisheries on pelagic redfish in the Irminger Sea and adjacent waters are generally regarded as having negligible impact on the habitat and other fish or invertebrate species due to very low bycatch and discard rates, characteristic of fisheries using pelagic gear.

### 21.7.6 Changes in the environment

The hydrography in the June/July 2013 survey show that temperature in the survey area is above average but it was lower than in 2011 in most of the surveyed area, except for the Irminger Current (ICES, 2013).

The increase of water temperature in the Irminger Sea may have an effect on spatial and vertical distribution of S. mentella in the feeding area (Pedchenko, 2005). The abundance and distribution of $S$. mentella in relation to oceanographic conditions were analysed in a special multistage workshop (WKREDOCE1-3). Based on 20 years of survey data, the results reveal the average relation of redfish to their physical habitat in shallow and intermediate waters: The most preferred latitude, longitude, depth, salinity and temperature for $S$. mentella are approximately $58^{\circ} \mathrm{N}, 40^{\circ} \mathrm{W}, 300 \mathrm{~m}, 34.89$ and $4.4^{\circ} \mathrm{C}$, respectively. The spatial distribution of S. mentella in the Irminger Sea mainly in waters $<500 \mathrm{~m}$ (and thus mainly relating to the "shallow" stock) appears strongly influenced by the Irminger Current Water (ICW) temperature changes, linked to the Subpolar Gyre (SPG) circulation and the North Atlantic Oscillation (NAO). The fish avoid waters mainly associated with the ICW $\left(>4.5^{\circ} \mathrm{C}\right.$ and $\left.>34.94\right)$ in the northeastern Irminger Sea, which may cause displacing towards the southwest, where fresher and colder water occurs (ICES 2012).
Results based on international redfish survey data suggest that the interannual distribution of fish above 500 m will shift in a southwest/northeast direction depending on integrated oceanographic conditions (ICES 2012).

### 21.8 References

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### 21.9 Tables

Table 21.2.1 Shallow Pelagic S. mentella (stock unit < 500 m ). Catches (in tonnes) by area as used by the Working Group.

| Year | VA | XII | XIV | NAFO 1 F | NAFO 2J | NAFO 2H | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 | 39783 | 20798 | 0 | 0 | 0 | 60581 |
| 1983 | 0 | 60079 | 155 | 0 | 0 | 0 | 60234 |
| 1984 | 0 | 60643 | 4189 | 0 | 0 | 0 | 64832 |
| 1985 | 0 | 17300 | 54371 | 0 | 0 | 0 | 71671 |
| 1986 | 0 | 24131 | 80976 | 0 | 0 | 0 | 105107 |
| 1987 | 0 | 2948 | 88221 | 0 | 0 | 0 | 91169 |
| 1988 | 0 | 9772 | 81647 | 0 | 0 | 0 | 91419 |
| 1989 | 0 | 17233 | 21551 | 0 | 0 | 0 | 38784 |
| 1990 | 0 | 7039 | 24477 | 385 | 0 | 0 | 31901 |
| 1991 | 0 | 9689 | 17048 | 458 | 0 | 0 | 27195 |
| 1992 | 106 | 22976 | 38709 | 0 | 0 | 0 | 62564 |
| 1993 | 0 | 66458 | 32500 | 0 | 0 | 0 | 100771 |
| 1994 | 665 | 77174 | 18679 | 0 | 0 | 0 | 96869 |
| 1995 | 77 | 78895 | 17895 | 0 | 0 | 0 | 100136 |
| 1996 | 16 | 22474 | 18566 | 0 | 0 | 0 | 41770 |
| 1997 | 321 | 18212 | 8245 | 0 | 0 | 0 | 27746 |
| 1998 | 284 | 21976 | 1598 | 0 | 0 | 0 | 24150 |
| 1999 | 165 | 23659 | 827 | 534 | 0 | 0 | 25512 |
| 2000 | 3375 | 17491 | 687 | 11052 | 0 | 0 | 33216 |
| 2001 | 228 | 32164 | 1151 | 5290 | 8 | 1751 | 41825 |
| 2002 | 10 | 24004 | 222 | 15702 | 0 | 3143 | 43216 |
| 2003 | 49 | 24211 | 134 | 26594 | 325 | 5377 | 56688 |
| 2004 | 10 | 7669 | 1051 | 20336 | 0 | 4778 | 33951 |
| 2005 | 0 | 6784 | 281 | 16260 | 5 | 4899 | 28229 |
| 2006 | 0 | 2094 | 94 | 12692 | 260 | 593 | 15734 |
| 2007 | 71 | 378 | 98 | 2843 | 175 | 2561 | 6126 |
| 2008 | 32 | 25 | 422 | 1580 | 0 | 0 | 2059 |
| 2009 | 0 | 210 | 2170 | 0 | 0 | 0 | 2380 |
| 2010 | 15 | 686 | 423 | 1074 | 0 | 0 | 2198 |
| 2011 | 0 | 0 | 234 | 0 | 0 | 0 | 234 |
| 2012 | 28 | 0 | 0 | 3113 | 32 | 0 | 3173 |
| 2013 | 32 | 13 | 40 | 1443 | 1 | 0 | 1529 |
| 2014 | 153 | 5068 | 489 | 713 | 0 | 0 | 6423 |
| 2015 | 161 | 2281 | 0 | 3119 | 34 | 0 | 5595 |

1982-1991 All pelagic catches assumed to be of the shallow pelagic stock
1992-1996 Guesstimates based on different sources (see text)
1997-2015 Catches from calculations based on jointed catch database and total landings

Table 21．2．2 Shallow pelagic S．mentella catches（in tonnes）in ICES Div．5．a，Subareas 12， 14 and NAFO Div．1F，2H and 2J by countries used by the Working Group．${ }^{*}$ Prior to 1991， the figures for Russia included Estonian，Latvian and Lithuanian catches．

|  |  |  | $\begin{aligned} & \stackrel{\nwarrow}{2} \\ & \text { D } \\ & \text { 岕 } \end{aligned}$ | $\begin{aligned} & \text { ひ } \\ & \text { O} \\ & \stackrel{\alpha}{4} \end{aligned}$ |  |  |  | $\begin{aligned} & \text { 号 } \\ & \underline{4} \\ & \underline{u} \end{aligned}$ | $\begin{aligned} & \frac{2}{d} \\ & \frac{\alpha}{4} \end{aligned}$ | $\sum_{\substack{4}}$ |  |  |  |  | $\begin{aligned} & 0 \\ & \text { 衣 } \\ & 0 \end{aligned}$ | ব S 0 0 0 | $\begin{aligned} & * \\ & \stackrel{\rightharpoonup}{4} \\ & \widehat{\sim} \\ & \underset{\sim}{n} \end{aligned}$ | $\begin{aligned} & \frac{z}{4} \\ & i \\ & i \end{aligned}$ | $\underset{J}{\text { v }}$ |  | ¢ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 581 |  | 60000 |  |  |  | 60581 |
| 1983 |  |  |  |  |  | 155 |  |  |  |  |  |  |  |  |  |  | 60079 |  |  |  | 60234 |
| 1984 | 2961 |  |  |  |  | 989 |  |  |  |  |  |  |  |  | 239 |  | 60643 |  |  |  | 64832 |
| 1985 | 5825 |  |  |  |  | 5438 |  |  |  |  |  |  |  |  | 135 |  | 60273 |  |  |  | 71671 |
| 1986 | 11385 |  |  | 5 |  | 8574 |  |  |  |  |  |  |  |  | 149 |  | 84994 |  |  |  | 105107 |
| 1987 | 12270 |  |  | 382 |  | 7023 |  |  |  |  |  |  |  |  | 25 |  | 71469 |  |  |  | 91169 |
| 1988 | 8455 |  |  | 1090 |  | 16848 |  |  |  |  |  |  |  |  |  |  | 65026 |  |  |  | 91419 |
| 1989 | 4546 |  |  | 226 |  | 6797 | 567 | 3816 |  |  |  |  |  |  | 112 |  | 22720 |  |  |  | 38784 |
| 1990 | 2690 |  |  |  |  | 7957 |  | 4537 |  |  |  |  |  | 7085 |  |  | 9632 |  |  |  | 31901 |
| 1991 |  |  | 2195 | 115 |  | 201 |  | 8724 |  |  |  |  |  | 6197 |  |  | 9747 |  |  |  | 27179 |
| 1992 | 628 |  | 1810 | 3765 | 2 | 6447 | 9 | 12080 |  | 780 | 6656 |  |  | 14654 |  |  | 15733 |  |  |  | 62564 |
| 1993 | 3216 |  | 6365 | 6812 |  | 16677 | 710 | 10167 |  | 6803 | 7899 |  |  | 14112 |  |  | 25229 |  |  | 2782 | 100771 |
| 1994 | 3600 |  | 17875 | 2896 | 606 | 15133 |  | 5897 |  | 13205 | 7404 |  |  | 6834 |  | 1510 | 16349 |  |  | 5561 | 96869 |
| 1995 | 2660 | 421 | 11798 | 3667 | 158 | 10714 | 277 | 8733 | 841 | 3502 | 16025 | 9 |  | 4288 |  | 2170 | 28314 | 1934 |  | 2230 | 100136 |
| 1996 | 1846 | 343 | 3741 | 2523 |  | 5696 | 1866 | 5760 | 219 | 572 | 5618 |  |  | 1681 |  | 476 | 9348 | 1671 | 137 | 273 | 41770 |
| 1997 |  | 102 | 3405 | 3510 |  | 9276 |  | 4446 | 28 |  |  |  |  | 330 | 776 | 367 | 3693 | 1812 |  |  | 27746 |
| 1998 |  |  | 3892 | 2990 |  | 9679 | 1161 | 1983 | 30 |  | 1734 |  |  | 701 | 12 | 60 | 89 | 1819 |  |  | 24150 |
| 1999 |  |  | 2055 | 1190 |  | 8271 | 998 | 3662 |  |  |  |  |  | 2098 | 6 | 62 | 6538 | 447 | 183 |  | 25512 |
| 2000 |  |  | 4218 | 486 |  | 5672 | 956 | 3766 |  |  | 430 |  |  | 2124 |  | 37 | 14373 | 1154 |  |  | 33216 |
| 2001 |  |  | 9 | 4364 |  | 4755 | 1083 | 14745 |  |  | 8269 |  |  | 947 |  | 256 | 5964 | 1433 |  |  | 41825 |
| 2002 |  |  |  | 719 |  | 5354 | 657 | 5229 |  | 1841 | 12052 |  |  | 1094 | 428 | 878 | 13958 | 1005 |  |  | 43216 |
| 2003 |  |  |  | 1955 |  | 3579 | 1047 | 4274 |  | 1269 | 21629 |  |  | 3214 | 917 | 1926 | 15418 | 1461 |  |  | 56688 |
| 2004 |  |  |  | 777 |  | 1126 | 750 | 5728 |  | 1114 | 3698 |  |  | 2721 | 1018 | 2133 | 13208 | 1679 |  |  | 33951 |
| 2005 |  |  |  | 210 |  | 1152 |  | 3086 |  | 919 | 1169 |  |  | 624 | 1170 | 2780 | 15562 | 1557 |  |  | 28229 |
| 2006 |  |  |  | 334 |  | 994 |  | 1293 |  | 1803 | 466 |  |  | 280 | 663 | 1372 | 4953 | 3576 |  |  | 15734 |
| 2007 |  |  | 209 | 98 |  | 0 |  | 71 |  | 186 | 467 |  |  |  | 189 | 529 | 4037 | 339 |  |  | 6126 |
| 2008 |  |  |  | 319 |  |  |  | 63 |  |  | 8 |  |  |  |  |  | 1597 | 73 |  |  | 2059 |
| 2009 |  |  |  | 87 |  |  |  | 5 |  |  | 138 |  |  |  |  |  | 649 | 1438 |  |  | 2380 |
| 2010 |  |  |  | 653 |  |  |  | 22 |  |  | 551 |  |  | 12 |  | 377 | 567 | 16 |  |  | 2198 |
| 2011 |  |  |  | 162 |  |  |  | 72 |  |  |  |  |  |  |  |  |  |  |  |  | 234 |
| 2012 |  |  |  |  |  |  |  | 28 |  |  |  |  |  |  |  |  | 3145 |  |  |  | 3173 |
| 2013 |  |  |  |  |  |  |  | 72 |  |  |  |  |  |  |  |  | 1457 |  |  |  | 1529 |
| 2014 |  |  |  |  |  |  |  | 355 |  |  | 287 |  |  |  |  |  | 5781 |  |  |  | 6423 |
| 2015 |  |  |  |  |  |  |  | 161 |  |  |  |  |  |  |  |  | 5434 |  |  |  | 5595 |

Table 21.2.3 Output from the GLM model used to standardize cpue

Call:
glm(formula $=$ lafli $\sim$ ltogtimi + factor(land $)+$ factor( yy ) + factor $(\mathrm{mm})+$ factor(skip), family = gaussian(), data = south)

Deviance Residuals:
Min 1Q Median 3Q Max
$-2.67560-0.274750 .015450 .282161 .70226$
Coefficients:
Estimate Std. Error $t$ value $\operatorname{Pr}(>|t|)$
(Intercept) $\quad 7.2883306000 .6215319011 .726398294 .487183 \mathrm{e}-27$ ltogtimi $\quad 1.0311890890 .0286543435 .987192171 .185172 \mathrm{e}-120$ factor(land)46 $0.3071080070 .196774191 .56071282 \quad 1.194800 \mathrm{e}-01$ factor(land) $58-0.6092223840 .59427534-1.02515171$ 3.059877e-01 factor(yy)1995 $-0.0145441450 .17246972-0.08432869$ 9.328425e-01 factor(yy)1996 $-0.5399670920 .20301506-2.659739058 .173648 \mathrm{e}-03$ factor(yy)1997 $-0.7810973750 .19187694-4.07082472$ 5.775636e-05 factor(yy)1998 $-0.5982056820 .20022972-2.98759682 \quad 3.006814 \mathrm{e}-03$ factor(yy)1999 -1.032123656 0.19849297-5.19979958 3.371986e-07 factor(yy)2000 $-0.4490670150 .18062595-2.48617105$ 1.337053e-02 factor(yy)2001 $-0.2940957490 .18731402-1.57006796 \quad 1.172876 \mathrm{e}-01$ factor(yy)2002 $-0.5534226980 .20779476-2.663314038 .089018 \mathrm{e}-03$ factor(yy)2003 $-0.4485304620 .20695582-2.16727635 \quad 3.087629 \mathrm{e}-02$ factor(yy)2004 $-0.9404675620 .19921557-4.72085375 \quad 3.382253 \mathrm{e}-06$ factor(y) $2005-0.8742280870 .21534893-4.059588746 .047701 \mathrm{e}-05$ factor(yy)2006 $-0.7925136220 .23511568-3.370739078 .318962 \mathrm{e}-04$ factor $(\mathrm{mm}) 3 \quad 0.4035399150 .62653390 \quad 0.64408313 \quad 5.199363 \mathrm{e}-01$ factor $(\mathrm{mm}) 4 \quad 0.0808863360 .599655290 .134888058 .927766 \mathrm{e}-01$ factor $(\mathrm{mm}) 5 \quad 0.6972894820 .59729418$ 1.16741383 2.438246e-01 factor $(\mathrm{mm}) 6 \quad 0.1065815040 .595821120 .178881728 .581323 \mathrm{e}-01$ factor $(\mathrm{mm}) 7 \quad 0.1560065390 .59913389 \quad 0.26038677$ 7.947160e-01 factor $(\mathrm{mm}) 8 \quad 0.2886879020 .60200469 \quad 0.47954427 \quad 6.318459 \mathrm{e}-01$ factor $(\mathrm{mm})^{9} \quad 0.1473727450 .603507550 .24419370 \quad 8.072215 \mathrm{e}-01$ factor(mm)10 $-0.0731373960 .61289180-0.119331669 .050799 \mathrm{e}-01$ factor $(\mathrm{mm}) 11 \quad-0.1114296360 .62872288-0.17723172$ 8.594272e-01 factor $(\mathrm{mm}) 12 \quad-0.6872076540 .84232729-0.815843994 .151349 \mathrm{e}-01$ factor(skip)118 -0.309179778 0.22143007-1.39628629 1.634983e-01 factor(skip)1270 $0.0376031490 .448280910 .08388300 \quad 9.331966 \mathrm{e}-01$ factor(skip)1273-0.628141253 0.22041607-2.84979787 4.629299e-03 factor(skip) 1279 -1.173362444 0.44513557-2.63596647 8.756942e-03 factor(skip) 1308 - $0.2669192650 .22303502-1.19675943 \quad 2.321967 \mathrm{e}-01$ factor(skip) $1328-0.2716542510 .21750992-1.24892811 \quad 2.125120 \mathrm{e}-01$ factor(skip) $1345-0.3894322550 .27300563-1.42646238 \quad 1.546113 \mathrm{e}-01$ factor(skip) $1351-0.2109225670 .30230014-0.697725674 .858042 \mathrm{e}-01$ factor(skip) $1360-0.1603370350 .37131520-0.431808436 .661421 \mathrm{e}-01$ factor(skip) $1365-0.0377783730 .28528994-0.132420988 .947261 \mathrm{e}-01$ factor(skip)1369 $0.0082218780 .232228210 .035404309 .717772 \mathrm{e}-01$ factor(skip) $1376-0.0793396290 .21104413-0.37593857$ 7.071865e-01 factor(skip)1408-0.360954071 $0.46295849-0.77966833$ 4.361041e-01 factor(skip)1412 -0.186735060 $0.60272438-0.30981833$ 7.568804e-01 factor(skip)1459-0.659207386 0.22905256-2.87797434 4.243932e-03 factor(skip) $1471-0.0677794360 .39810737-0.17025416 \quad 8.649070 \mathrm{e}-01$ factor(skip)1472 -0.243213212 $0.33706786-0.72155563$ 4.710413e-01 factor(skip)1473-0.831933012 0.45025953-1.84767443 6.547885e-02 factor(skip) $1552-1.3085858940 .61116338-2.14113925$ 3.294138e-02 factor(skip) 1578 -1.486687432 0.38045634-3.90764269 1.115534e-04 factor(skip) $1579-0.4747097490 .30501933-1.55632678 \quad 1.205189 \mathrm{e}-01$ factor(skip)1585-0.553949127 0.61783175-0.89660191 3.705373e-01 factor(skip)1628 0.0488619840 .452916860 .10788290 9.141494e-01 factor(skip) $180-0.5326137340 .18564922-2.868925304 .364387 \mathrm{e}-03$ factor(skip)1833-0.296067754 0.22785023-1.29939633 1.946488e-01
factor(skip)1868-0.104954736 0.22921245-0.45789282 6.473088e-01 factor(skip) $18800.0041530550 .258263610 .016080689 .871790 \mathrm{e}-01$ factor(skip) $19020.2040439870 .284172820 .718027824 .732111 \mathrm{e}-01$ factor(skip) 1976-0.380940434 $0.61538320-0.619029635 .362928 \mathrm{e}-01$ factor(skip) 1977-0.774106835 0.33815309-2.28922009 2.265145e-02 factor(skip)2165 $0.1050475900 .205808960 .510413116 .100784 \mathrm{e}-01$ factor(skip)2170-0.122213348 0.20408250-0.59884286 5.496585e-01 factor(skip)2182-0.454140930 $0.23283220-1.95050737$ 5.190006e-02 factor(skip)2184-0.295249414 0.25222782-1.17056639 $2.425561 \mathrm{e}-01$ factor(skip)2203-0.136558045 0.20059787-0.68075523 4.964689e-01 factor(skip)2212 $0.1833021430 .304962760 .601064025 .481798 \mathrm{e}-01$ factor(skip)2236-0.581565095 0.26502996-2.19433717 2.885678e-02 factor(skip)2265 $0.2397188650 .439515190 .545416575 .858086 \mathrm{e}-01$ factor(skip)2592-0.282434578 $0.59801605-0.472285956 .370121 \mathrm{e}-01$ factor(skip)3033-0.283499142 $0.72458991-0.391254616 .958431 \mathrm{e}-01$ factor(skip)3135-0.016478186 $0.66105345-0.02492716$ 9.801270e-01 factor(skip) 3156-0.260805362 $0.61679034-0.422842816 .726652 \mathrm{e}-01$ factor(skip) $3382-0.4234249190 .62159313-0.681193054 .961922 \mathrm{e}-01$ factor(skip)3523-0.395258535 0.72563919-0.54470396 5.862982e-01 factor(skip)3542 $0.0183557450 .619941890 .029608829 .763956 \mathrm{e}-01$ factor(skip)3709-0.609676578 $0.64465767-0.94573695$ 3.449242e-01 factor(skip)934-1.054646713 0.17107235-6.16491621 1.912436e-09--Signif. codes: $0^{\prime * * * '} 0.001^{\prime * * \prime} 0.01^{\prime * \prime} 0.05^{\text {'.' }} 0.1^{\prime}$ ' 1
(Dispersion parameter for gaussian family taken to be 0.3450127 )

Null deviance: 989.53 on 458 degrees of freedom
Residual deviance: 131.45 on 381 degrees of freedom AIC: 886.64

```
Number of Fisher Scoring iterations: }
Analysis of Deviance Table
Model: gaussian, link: identity
Response: lafli
Terms added sequentially (first to last)
Df Deviance Resid. DfResid. Dev F Pr
NULL 428 934.30
ltogtimi 1 682.16 427 252.14 2126.3228<2.2e-16***
factor(land) 2 38.99 425 213.15 60.7682<2.2e-16***
factor(yy) 12 43.18 413 169.96 11.2167<2.2e-16 ***
factor(mm) 10}1017.04 403 152.92 5.3122 2.600e-07 ****
factor(skip) 47 38.71 356 114.21 2.5673 5.376e-07 ***
Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
```

Table 21.6.1 Shallow Pelagic S. mentella. Results for the acoustic survey indices 1991-2013 from shallower than the scattering layer, trawl estimates within the deep scattering layer and shallower than 500 m , and area coverage of the survey in the Irminger Sea and adjacent waters. No estimates are available for 2015.

| Area covered (1,000 |  |  | Trawl estimates (1,000 T) |
| :---: | :---: | :---: | :---: |
| Year | NM2) | Acoustic estimates (1,000 T) |  |
| 1991 | 105 | 2,235 |  |
| 1992 | 190 | 2,165 |  |
| 1993 | 121 | 2,556 |  |
| 1994 | 190 | 2,190 |  |
| 1995 | 168 | 2,481 |  |
| 1996 | 253 | 1,576 |  |
| 1997 | 158 | 1,225 |  |
| 1999 | 296 | 614 |  |
| 2001 | 420 | 716 | 565 |
| 2003* | 405 | 89* | 92* |
| 2005 | 386 | 550 | 392 |
| 2007 | 349 | 372 | 283 |
| 2009 | 360 | 108 | 331 |
| 2011 | 343 | 123 | 361 |
| 2013 | 340 | 91 | 200 |

* The 2003 biomass estimate is considered as inconsistent as the survey was carried out about one month earlier than usual, and a marked seasonal effect was observed.

Table 21.6.2. Results (biomass in ' 000 t ) for the international surveys conducted since 1994, for redfish shallower than the DSL for each subarea (see Figure 21.6.5 for area definition) and total.

| SUB-AREA |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | A | B | C | D | E | F | TOTAL |
| 1994 | 673 | 1,228 | - | 63 | 226 |  | 2,190 |
| 1996 | 639 | 749 | - | 33 | 155 |  | 1,576 |
| 1999 | 72 | 317 | 16 | 42 | 167 |  | 614 |
| 2001 | 88 | 220 | 30 | 267 | 103 | 7 | 716 |
| 2003 | 32 | 46 | 1 | 2 | 10 | 0 | 89 |
| 2005 | 121 | 123 | 0 | 87 | 204 | 17 | 551 |
| 2007 | 80 | 95 | 0 | 53 | 142 | 3 | 372 |
| 2009 | 39 | 48 | 4 | 1 | 15 | 1 | 108 |
| 2011 | 5 | 74 | 0 | 3 | 40 | 1 | 123 |
| 2013 | 9 | 33 | 2 | 5 | 42 | 0 | 91 |

### 21.10Figures



Figure 21.2.1 Fishing areas and total catch of pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters 1989-2012. Data are from the Faroe Islands (1995-2012), Iceland (1989-2012) and Norway (1992-2003). The catches in the legend are given as tonnes per square nautical mile. The blue box represents the management unit for the northern fishing area.


Figure 21.2.1 (Cont.) Fishing areas and total catch of pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters 1989-2012. Data are from the Faroe Islands (1995-2012), Iceland (1989-2012) and Norway (1992-2003). The catches in the legend are given as tonnes per square nautical mile. The blue box represents the management unit for the northern fishing area.


Figure 21.2.2 Landings of shallow pelagic S. mentella (Working Group estimates, see Table 21.2.1).


Figure 21.2.3 Trends in standardized cpue of the shallow pelagic S. mentella fishery in the Irminger Sea and adjacent waters, based on logbook data from Faroes, Iceland, Norway, and Greenland.


Figure 21.2.4 Residuals from the GLM model used to standardize cpue, based on logbook data from Faroe Islands, Iceland, Greenland and Norway.


Figure 21.3.1 Length distribution from Icelandic landings of shallow pelagic S. mentella.


Figure 21.6.1 Pelagic S. mentella. Acoustic estimates (average sa values by 5 NM sailed) shallower than the deep-scattering layer (DSL) from the joint trawl-acoustic survey in June/July 2013.


Figure 21.6.2. Redfish acoustic estimates shallower than the DSL. Average sA values within statistical rectangles during the joint international redfish survey in June/July 2013.


Figure 21.6.3 Redfish trawl estimates within the DSL shallower than 500 m (type $\mathbf{2}$ trawls). sA values calculated by the trawl method (chapter 2.2.3) during the joint international redfish survey in June/July 2013.


Figure 21.6.4. Overview of acoustic survey indices (thousand tonnes) from above the scattering layer (red filled circle), trawl estimates within the scattering layer and shallower than 500 m (black triangle), and aerial coverage (NM2) of the survey (black open circle) in the Irminger Sea and adjacent waters.


Figure 21.6.5 Sub-areas A-F used on international surveys for redfish in the Irminger Sea and adjacent waters, and divisions for biological data (Northeast, Southwest and Southeast; boundaries marked by broken lines).


Figure 21.6.6 Length distribution of redfish in the trawls, by geographical areas and total, from fish caught shallower than 500 m (in 2013).

## 22 Deep Pelagic Sebastes mentella

### 22.1 Stock description and management unit

This section addresses the fishery for the biological stock deep pelagic S. mentella in the Irminger Sea and adjacent areas: NAFO 1-2, ICES 5, 12, and 14 at depths $>500 \mathrm{~m}$, including demersal habitats west of the Faroe Islands. This stock corresponds to the management unit in the northeast Irminger Sea (ICES areas 5.a, 12 and 14).

The following text table summarizes the available information from fishing fleets in the Irminger Sea and adjacent waters in 2015. It should be noted that some these fleets are also fishing the Shallow Pelagic stock:

|  | COUNTRY |
| :--- | :--- |
| Faroes | NUMBER OF TRAWLERS |
| Iceland | 2 factory trawlers |
| Germany | 5 factory trawlers |
| Lithuania | 1 factory trawler |
| Norway | 1 factory trawler |
| Portugal | 3 factory trawlers |
| Russia | 1 factory trawler |
| Spain | 16 factory trawlers |

### 22.2 The fishery

The historic development of the fishery can be found in the Stock Annex. Tables 22.2.1 and 22.2.2 show annual catches, as estimated by the Working Group, disaggregated by ICES and NAFO regulatory areas and by country, respectively.

The changes in the spatial pattern of the fishery for the period 1992-2015 are shown in Figure 22.2.1, and annual catches are presented in Figure 22.2.2. Catches increased by 4000 t in 2015 to 27433 t (Table 22.2.2).

Standardized cpue series for Faroe Islands, Iceland, Greenland, and Norway 1994-2015 are estimated with a GLM model including the factors year, ship, month and towing time. The results from the model show that the cpue oscillates without trend since 1995 (Figure 22.2.3). The model output is shown in Table 21.2 .3 and the residuals are in Figure 22.2.4. The cpue index increased from about 0.3 in 2012 to $>1.0$ in 2013, but decreased again to the lowest level in 2014 and 2015.

### 22.3 Biological information

The length distribution from Icelandic landings for the period 1991-2015 is shown in Figure 22.3.1. Mean length ranged from 40.05 to 41.17 cm all years but 2002 and 2006, but since 2007 has oscillated between 37.88 and 39.43 cm .

### 22.4 Discards

Discards are not considered to be significant for the time being, according to available data from various institutes.

### 22.5 Illegal, Unregulated and Unreported Fishing (IUU)

The Group had again difficulties in obtaining catch estimates from several fleets. Furthermore, there are problems caused by misreported catches. The Group requests NEAFC and NAFO to provide ICES in time with all the necessary information.

### 22.6 Surveys

The last international trawl-acoustic survey took place in 2015 and it is described in detail in ICES CM WGRS REPORT 2015 (ICES, 2015). The survey was carried out by Iceland and Germany. The participation of Russia was cancelled at the beginning of May 2015 because of reasons not specified. For this reason the scope of the survey had to be altered and the emphasis was on covering the deep pelagic stock found below 500 m .

### 22.6.1 Survey trawl estimates

Considering the conclusion of WKREDS (ICES, 2009a) and the recommendation of ICES on stock structure of redfish in the Irminger Sea and adjacent waters, the Group decided in the planning meeting (ICES, 2009b) to sample redfish separately above and below 500 m , i.e. to sample redfish as was done in the 1999, 2001 and 2003 surveys . The deep identification hauls covered the depth layers (headline) $550 \mathrm{~m}, 700 \mathrm{~m}$, and 850 m .

The most recent trawl-acoustic survey on pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters was carried out by Iceland and Germany in June/ July 2015. Approximately $200000 \mathrm{NM}^{2}$ were covered. A total biomass of 196000 t was estimated, significantly below the 280000 t of 2013 (Table 22.6.2). The results showed large biomass declines in subareas A and B, the main distribution area of the stock (see Figure 22.6.1 for area definition) (Table 21.6.2). Biological samples from the trawls taken at depth $>500 \mathrm{~m}$ showed a mean length of 38.6 cm , which is similar as the mean length in 2013. Figure 22.6 . 2 shows the spatial distribution of samples used in the survey and Figure 22.6 .3 shows the corresponding length distribution.

### 22.7 Methods

The assessment of pelagic redfish in the Irminger Sea and adjacent waters is based on survey indices, catches, cpue and biological data.

### 22.8 Reference points

For pelagic redfish in the Irminger Sea and adjacent waters, no analytical assessment is carried out due to data uncertainties and the lack of reliable age data. Thus, no reference points can be derived.

### 22.9 State of the stock

### 22.9.1 Short-term forecast

For pelagic redfish in the Irminger Sea and adjacent waters, no analytical assessment is being carried out due to data uncertainties and the lack of reliable age data. Thus, no short-term forecasts can be derived.

### 22.9.2 Uncertainties in assessment and forecast

### 22.9.2.1 Data considerations

Preliminary official landings data were provided by the ICES Secretariat, NEAFC and NAFO, and various national data were reported to the Group. The Group, however, repeatedly faces problems to obtain reliable catch data due to unreported catches of pelagic redfish and lack of catch data disaggregated by depth from some countries.

As in previous years, detailed descriptions on the horizontal, vertical and seasonal distribution of the fisheries are given.

The need and importance of having catch and biological data disaggregated by depth from all nations taking part in the fishery cannot be stressed strongly enough, and the Group urges all nations involved on supplying better data. With this need in mind, ICES sent a data call to all EU countries participating in the redfish fishery, encouraging stockholders to deliver detailed catch data before the WG would meet, but the response was very limited.

### 22.9.2.2 Assessment quality

The results of the international trawl-acoustic survey are given in section 21.6. Given the high variability of the correlation between trawl and acoustic estimates as well as the assumptions that need to be made about constant catchability across depth and areas, the uncertainty of these estimates is very high.

It is not known to what extent cpue reflect changes in the stock status of pelagic $S$. mentella, since the fishery focuses on aggregations. Therefore, stable or increasing cpue series might not indicate or reflect actual trends in stock size, although decreasing cpue indices are likely to reflect a decreasing stock.

### 22.9.3 Comparison with previous assessment and forecast

The data available for evaluating the stock status are similar to last year.

### 22.9.4 Management considerations

The Group needs more and better data and requests that NEAFC and NAFO provide ICES with all information leading to more reliable catch statistics.

The main feature of the fishery since 1998 is a clear distinction between two widely separated fishing grounds with pelagic redfish fished at different seasons and different depths. Since 2000, the southwestern fishing grounds extended also into the NAFO Convention Area. Biological data, however, suggest that the aggregations in the NAFO Convention Area do not constitute a separate stock. The NAFO Scientific Council agreed with this conclusion (NAFO, 2005). The Group concludes that at this time there is not enough scientific basis available to propose an appropriate split of the total TAC among the two fisheries/areas.

The 9500 t TAC set by NEAFC for 2015 was overshot by about 18000 t . This excess is due to the unilateral decision of the Russian Federation to self-allocate an annual TAC, which was 27300 t for 2015. It was taken from both Shallow and Deep pelagic (20 214 t) stocks, since the Russian Federation does not agree on the division of the S. mentella management units.

### 22.9.5 Ecosystem considerations

The fisheries on pelagic redfish in the Irminger Sea and adjacent waters are generally regarded as having negligible impact on the habitat and other fish or invertebrate species due to very low bycatch and discard rates, characteristic of fisheries using pelagic gear.

### 22.9.6 Changes in the environment

The hydrography in the survey of June/July 2013 show that temperature in the survey area is above average but it was lower than in 2011 in most of the surveyed area, except for the Irminger Current (ICES, 2013).
The increase of water temperature in the Irminger Sea may have an effect on spatial and vertical distribution of S. mentella in the feeding area (Pedchenko, 2005). The abundance and distribution of $S$. mentella in relation to oceanographic conditions were analysed in a special multistage workshop (WKREDOCE1-3, see ICES 2012b). Based on 20 years of survey data, the results reveal the average relation of redfish to their physical habitat in shallow and intermediate waters: The most preferred latitude, longitude, depth, salinity and temperature for $S$. mentella are approximately $58^{\circ} \mathrm{N}, 40^{\circ} \mathrm{W}, 300 \mathrm{~m}$, 34.89 and $4.4^{\circ} \mathrm{C}$, respectively. The spatial distribution of $S$. mentella in the Irminger Sea mainly in waters $<500 \mathrm{~m}$ (and thus mainly relating to the "shallow" stock) appears strongly influenced by the Irminger Current Water (ICW) temperature changes, linked to the Subpolar Gyre (SPG) circulation and the North Atlantic Oscillation (NAO). The fish avoid waters mainly associated with the ICW ( $>4.5^{\circ} \mathrm{C}$ and salinity $>34.94$ ) in the northeastern Irminger Sea, which may cause displacing towards the southwest, where fresher and colder water occurs (ICES 2012b).

Results based on international redfish survey data suggest that the interannual distribution of fish above 500 m will shift in a southwest/northeast direction depending on integrated oceanographic conditions (ICES 2012). Whether the results of the study mentioned are applicable to the conditions for the deep pelagic stock needs further investigation.

### 22.10 WKDEEPRED 2016

The Workshop on Assessment and Catch Advice for Deep Pelagic Redfish in the Irminger Sea (WKDEEPRED) will be held in 23-25 August 2016 in ICES HQ, Copenhagen, Denmark. The workshop will:
a ) Agree on the most appropriate approach for the assessment of this stock;
b ) If an analytical assessment is agreed for the stock (categories 1 or 2), estimate Biological Reference Points and agree on the short-term forecast method. Otherwise, consider proxy reference points to the extent possible and agree on a methodology for catch advice for the stock consistent with ICES advisory rules;
c) Document the agreed methodology in a stock annex;
d ) In response to NEAFCs request to ICES "to review the approach that serves as basis for the catch advice for the deep pelagic redfish stock in the Irminger Sea to ensure conformity with ICES advisory rules", conduct a stock assessment and prepare catch advice for the stock for 2017 in accordance with a) and $b$ ).

### 22.11References

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### 22.12Tables

Table 22.2.1 Deep Pelagic S. mentella (stock unit > 500 m ). Catches (in tonnes) by area as used by the Working Group.

| Year | VA | XII | XIV | NAFO 1F | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 0 | 7 | 52 | 0 | 59 |
| 1992 | 1862 | 280 | 1257 | 0 | 3398 |
| 1993 | 2603 | 6068 | 6393 | 0 | 15064 |
| 1994 | 14807 | 16977 | 20036 | 0 | 51820 |
| 1995 | 1466 | 53141 | 21100 | 0 | 75707 |
| 1996 | 4728 | 20060 | 113765 | 0 | 138552 |
| 1997 | 14980 | 1615 | 78485 | 0 | 95079 |
| 1998 | 40328 | 444 | 52046 | 0 | 92818 |
| 1999 | 36359 | 373 | 47421 | 0 | 84153 |
| 2000 | 41302 | 0 | 51811 | 0 | 93113 |
| 2001 | 27920 | 0 | 59073 | 0 | 86993 |
| 2002 | 37269 | 2 | 65858 | 0 | 103128 |
| 2003 | 46627 | 21 | 57648 | 0 | 104296 |
| 2004 | 14446 | 0 | 77508 | 0 | 91954 |
| 2005 | 11726 | 0 | 33759 | 0 | 45485 |
| 2006 | 16452 | 51 | 50531 | 254 | 67288 |
| 2007 | 17769 | 0 | 40748 | 0 | 58516 |
| 2008 | 4602 | 0 | 25443 | 0 | 30045 |
| 2009 | 16828 | 4658 | 32920 | 0 | 54406 |
| 2010 | 8552 | 0 | 50736 | 0 | 59288 |
| 2011 | 0 | 7 | 47326 | 0 | 47333 |
| 2012 | 5530 | 608 | 26668 | 0 | 32806 |
| 2013 | 5274 | 0 | 40778 | 0 | 46052 |
| 2014 | 603 | 0 | 23152 | 0 | 23755 |
| 2015 | 1821 | 0 | 25612 | 0 | 27433 |

Table 22．2．2．Deep pelagic S．mentella catches（in tonnes）in ICES Divisions 5．a，Subareas 12,14 and NAFO Div．1F，2H and 2J by countries used by the Working Group．

|  |  | $\begin{aligned} & \text { à } \\ & \frac{4}{4} \\ & \frac{2}{4} \end{aligned}$ | $\begin{aligned} & \stackrel{\checkmark}{2} \\ & 0 \\ & \stackrel{0}{4} \end{aligned}$ |  |  |  |  |  | $\begin{aligned} & \text { z } \\ & \frac{\alpha}{4} \end{aligned}$ | $\sum_{\vdots}^{\leftrightarrows}$ |  | $\begin{aligned} & \text { 号 } \\ & \text { d } \\ & \text { 岂 } \\ & \text { 己 } \end{aligned}$ |  | $\begin{aligned} & 0 \\ & \frac{2}{4} \\ & 0 \end{aligned}$ |  | $\begin{aligned} & \stackrel{\checkmark}{n} \\ & \stackrel{y}{x} \end{aligned}$ | $\begin{aligned} & \frac{2}{1} \\ & i \\ & i \end{aligned}$ | $\underset{\sim}{\square}$ | $$ | $\begin{aligned} & \stackrel{1}{5} \\ & \stackrel{\rightharpoonup}{\circ} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |


| 1991 |  |  |  |  |  |  |  | 59 |  |  |  |  |  |  |  |  |  |  |  | 59 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 |  |  |  |  |  |  |  | 3398 |  |  |  |  |  |  |  |  |  |  |  | 3398 |
| 1993 |  |  |  | 310 |  | 1135 |  | 12741 |  |  |  |  | 878 |  |  |  |  |  |  | 15064 |
| 1994 |  |  |  |  |  | 2019 |  | 47435 |  |  |  |  | 523 |  | 377 | 1465 |  |  |  | 51820 |
| 1995 | 1140 | 181 | 5056 | 1572 | 68 | 8271 | 1579 | 25898 | 396 | 1501 | 6868 | 4 | 3169 |  | 2955 | 15868 | 227 |  | 956 | 75707 |
| 1996 | 1654 | 307 | 3351 | 3748 |  | 15549 | 1671 | 57143 | 196 | 512 | 5031 |  | 5161 |  | 1903 | 36400 | 5558 | 123 | 245 | 138552 |
| 1997 |  | 9 | 315 | 435 |  | 11200 |  | 36830 | 3 |  |  |  | 2849 |  | 3307 | 33237 | 6895 |  |  | 95079 |
| 1998 |  |  | 76 | 4484 |  | 8368 | 302 | 46537 | 1 |  | 34 |  | 438 |  | 4073 | 25748 | 2758 |  |  | 92818 |
| 1999 |  |  | 53 | 3466 |  | 8218 | 3271 | 40261 |  |  |  |  | 3337 |  | 4240 | 11419 | 9885 | 5 |  | 84153 |
| 2000 |  |  | 7733 | 2367 |  | 6827 | 3327 | 41466 |  |  | 0 |  | 3108 |  | 3694 | 14851 | 9740 |  |  | 93113 |
| 2001 |  |  | 878 | 3377 |  | 5914 | 2360 | 27727 |  |  | 7515 |  | 4275 |  | 2488 | 23810 | 8649 |  |  | 86993 |
| 2002 |  |  | 15 | 3664 |  | 7858 | 3442 | 39263 |  |  | 9771 |  | 4197 |  | 2208 | 25309 | 7402 |  |  | 103128 |
| 2003 |  |  |  | 3938 |  | 7028 | 3403 | 44620 |  |  | 0 |  | 5185 |  | 2109 | 28638 | 9374 |  |  | 104296 |
| 2004 |  |  |  | 4670 |  | 2251 | 2419 | 31098 |  |  | 0 |  | 6277 | 1889 | 2286 | 31067 | 9996 |  |  | 91954 |
| 2005 |  |  |  | 1800 |  | 1836 | 1431 | 12919 |  |  | 1027 |  | 3950 | 1240 | 1088 | 16323 | 3871 |  |  | 45485 |
| 2006 |  |  |  | 3498 |  | 1830 | 744 | 20942 |  |  | 1294 |  | 5968 | 1356 | 1313 | 23670 | 6673 |  |  | 67288 |
| 2007 |  |  |  | 2902 |  | 1110 | 1961 | 18097 |  | 575 | 1394 |  | 4628 | 636 | 2067 | 21337 | 3810 |  |  | 58516 |
| 2008 |  |  |  | 2632 |  |  | 1170 | 6723 |  |  | 749 |  | 571 | 219 | 1733 | 15106 | 1142 |  |  | 30045 |
| 2009 |  |  |  | 3206 |  |  | 1519 | 15125 |  | 1355 | 2613 |  |  | 178 | 1596 | 25309 | 2907 |  |  | 54006 |
| 2010 |  |  |  | 3195 |  |  | 1932 | 14772 |  | 1963 | 2228 |  | 2388 | 3 | 2203 | 22803 | 7801 |  |  | 59288 |
| 2011 |  |  |  | 2028 |  | 1787 |  | 11994 |  | 845 | 1348 |  | 1066 |  | 1540 | 22364 | 4361 |  |  | 47333 |
| 2012 |  |  |  | 1438 |  | 1523 |  | 5912 |  | 724 | 558 |  | 3362 |  | 250 | 18377 | 632 |  |  | 32806 |
| 2013 |  |  |  | 1882 |  | 1176 |  | 8545 |  | 1200 | 1163 |  | 2979 |  |  | 26463 | 2644 |  |  | 46052 |
| 2014 |  |  |  | 721 |  | 890 |  | 2081 |  | 867 | 1024 |  | 1965 |  |  | 15475 | 732 |  |  | 23755 |
| 2015 ${ }^{1)}$ |  |  |  | 779 |  | 918 |  | 1968 |  |  | 330 |  | 1547 |  | 202 | 20214 | 1475 |  |  | 27433 |

1）Provisional．Official Spanish catch data were lower than the data provided by NEAFC and the WG decided to use the highest catch data as a precautionary measure．

Table 22.2.3 Output from the GLM model used to standardize cpue - NOT UPDATED
Call:
glm(formula $=$ lafli $\sim$ ltogtimi + factor(land $)+$ factor $(y y)+$ factor $(\mathrm{mm})+$ factor(skip), family $=$ gaussian () , data $=$ north $)$

Deviance Residuals:
Min 1Q Median 3Q Max
$\begin{array}{llllll}-3.5126 & -0.2410 & 0.0168 & 0.2924 & 1.4568\end{array}$

Coefficients: (3 not defined because of singularities)
Estimate Std. Error $t$ value $\operatorname{Pr}(>|t|)$
(Intercept) $7.8715952790 .3909419420 .1349469382 .522077 \mathrm{e}-78$ Itogtimi $1.0514389930 .0169557262 .0108867870 .000000 \mathrm{e}+00$ factor(land)6 $-0.1987745460 .35054379-0.5670462565 .707845 \mathrm{e}-01$ factor(land)46 $0.3564163710 .13062029 \quad 2.7286447886 .448602 \mathrm{e}-03$ factor(land)58 $0.3859056630 .35297372 \quad 1.093298568$ 2.744732e-01 factor(land)69 $0.1314272160 .214801140 .6118552955 .407447 \mathrm{e}-01$ factor(yy)1995 $-0.5444782240 .09104861-5.9800829242 .906186 \mathrm{e}-09$ factor(yy)1996 $-0.5913507580 .08559433-6.9087610037 .764598 \mathrm{e}-12$ factor(yy)1997 -1.074579737 0.08522551-12.608663463 2.120951e-34 factor(yy)1998 $-0.6956380420 .08486144-8.1973398116 .028342 \mathrm{e}-16$ factor(yy)1999 $-0.7979159670 .08474619-9.4153611452 .186144 \mathrm{e}-20$ factor(yy)2000 $-0.4317900780 .08594177-5.0242164755 .788362 \mathrm{e}-07$ factor(yy)2001 -0.955769159 0.08486379-11.262391106 4.296930e-28 factor(yy)2002 $-0.5755450270 .08596398-6.6951883723 .244414 \mathrm{e}-11$ factor(yy)2003 $-0.3162932040 .08682628-3.6428280982 .806914 \mathrm{e}-04$ factor(yy)2004 -1.016098316 $0.08870892-11.454296393$ 5.867807e-29 factor(yy)2005 -1.325075546 0.09344673-14.180009215 1.955001e-42 factor(yy)2006 $-0.9199058300 .09688906-9.4944241991 .079864 \mathrm{e}-20$ factor(yy)2007 $-0.6814049910 .10068657-6.7675859402 .007085 \mathrm{e}-11$ factor(yy)2008 $-1.0392929730 .11665575-8.9090594271 .776186 \mathrm{e}-18$ factor(yy)2009 $-0.5758115150 .10378067-5.5483503803 .516366 \mathrm{e}-08$ factor(yy)2010 $-0.3373305720 .10886481-3.0986189921 .987589 \mathrm{e}-03$ factor(yy)2011 $-0.7157143570 .10794155-6.6305735864 .961323 \mathrm{e}-11$ factor(yy)2012 -1.309676579 0.11568546-11.321012872 2.345629e-28 factor(yy)2013 $0.0088683250 .152009230 .0583407009 .534866 \mathrm{e}-01$ factor $(\mathrm{mm}) 3 \quad-0.8126598030 .39174110-2.0744818623 .823868 \mathrm{e}-02$ factor $(\mathrm{mm}) 4 \quad-0.3785501090 .37671843-1.0048622013 .151575 \mathrm{e}-01$ factor $(\mathrm{mm}) 5 \quad-0.1804825970 .37822841-0.4771788536 .333181 \mathrm{e}-01$ factor(mm)6 $\quad-0.3330639150 .37796344-0.8812066923 .783752 \mathrm{e}-01$ factor $(\mathrm{mm}) 7 \quad-0.5038716480 .37798537-1.3330453631 .827596 \mathrm{e}-01$ factor $(\mathrm{mm}) 8 \quad-0.6088381370 .38197175-1.593934978$ 1.112032e-01 factor $(\mathrm{mm})^{9} \quad-0.4593266970 .39365610-1.1668222632 .435045 \mathrm{e}-01$ factor $(\mathrm{mm}) 10 \quad-0.7581709300 .43452708-1.7448185828 .126201 \mathrm{e}-02$ factor(mm)11 $-0.7468334340 .50556889-1.4772139941 .398700 \mathrm{e}-01$ factor(skip)118 -0.267350277 $0.14387130-1.8582600046 .336686 \mathrm{e}-02$ factor(skip)1265-0.305357861 $0.22335074-1.3671674611 .718184 \mathrm{e}-01$ factor(skip)1268-0.266852354 $0.48481836-0.5504171805 .821315 \mathrm{e}-01$ factor(skip)1270 -0.152220760 $0.13168835-1.1559166532 .479360 \mathrm{e}-01$ factor(skip)1273-0.419400828 $0.13225616-3.1711250081 .555377 \mathrm{e}-03$ factor(skip)1279-0.447449821 0.22139223-2.021072797 4.348488e-02 factor(skip)1308-0.038459005 0.12407714 -0.309960450 7.566427e-01 factor(skip)1328-0.144677028 $0.13096482-1.1047014502 .695014 \mathrm{e}-01$ factor(skip)1345-0.458372216 $0.12962963-3.5360141154 .210561 \mathrm{e}-04$ factor(skip)1351-0.357088024 0.13798946-2.587792070 9.771199e-03 factor(skip) $1360-0.1152608780 .13128164-0.8779664893 .801305 \mathrm{e}-01$ factor(skip)1365-0.295965760 $0.14818615-1.9972566104 .601360 \mathrm{e}-02$ factor(skip)1369-0.011320574 0.14158965 -0.079953400 9.362871e-01 factor(skip)1376-0.169473795 $0.12691871-1.3352940731 .820230 \mathrm{e}-01$ factor(skip)1395-0.409995924 0.26206284-1.564494724 1.179543e-01 factor(skip) $1408-1.1017738150 .48734615-2.2607623082 .394521 \mathrm{e}-02$ factor(skip)1412-0.111357948 0.29537612 -0.377003899 7.062346e-01
factor(skip)1459-0.583761981 0.13242885 -4.408117879 1.132118e-05 factor(skip)1471-0.613726934 0.17679359-3.471432056 5.353333e-04 factor(skip)1472-0.511423665 0.16493944 -3.100675325 1.973945e-03 factor(skip)1473-0.952361655 0.21265005 -4.478539431 8.201613e-06 factor(skip)1484-1.433135836 0.48533948 -2.952852375 3.207303e-03 factor(skip)1497-1.418776803 $0.35333866-4.0153454776 .288939 \mathrm{e}-05$ factor(skip)1530 -0.740040738 $0.48506624-1.5256488311 .273501 \mathrm{e}-01$ factor(skip)1536-1.814090714 0.48651089-3.728777180 2.010160e-04 factor(skip)1552-1.430344940 $0.29607192-4.8310725841 .525962 \mathrm{e}-06$ factor(skip)1553-0.003402939 $0.35390074-0.009615518$ 9.923296e-01 factor(skip)1578-0.354319033 $0.15025452-2.3581255911 .852067 \mathrm{e}-02$ factor(skip)1579-0.071762545 0.12696644 -0.565208749 5.720331e-01 factor(skip)1585-0.417372263 $0.17163124-2.4317965211 .516373 \mathrm{e}-02$ factor(skip)1628-1.097387934 0.29532540 -3.715860332 2.114507e-04 factor(skip) $180-0.5226640610 .11197449-4.6677065743 .373762 \mathrm{e}-06$ factor(skip)1833-0.025339243 0.12377825 -0.204714826 8.378282e-01 factor(skip)1868-0.141108499 $0.12362729-1.1414024702 .539209 \mathrm{e}-01$ factor(skip) $1880-0.3005382110 .13797726-2.1781720892 .957946 \mathrm{e}-02$ factor(skip)1902-0.122037774 $0.13376781-0.912310478$ 3.617811e-01 factor(skip)1903-0.503244254 0.29824691 -1.687341029 9.178708e-02 factor(skip)1976-0.628200177 0.20114205 -3.123166888 1.830228e-03 factor(skip)1977-0.169902957 0.14830443 -1.145636427 $2.521647 \mathrm{e}-01$ factor(skip)2107-0.796278905 0.35043313 -2.272270619 2.323928e-02 factor(skip)2165 $0.0724955880 .13389141 \quad 0.5414506385 .882934 \mathrm{e}-01$ factor(skip)2170 -0.090472494 0.12251178 -0.738479960 4.603614e-01 factor(skip)2182-0.227474427 $0.12989539-1.751212416$ 8.015439e-02 factor(skip)2184-0.083675892 $0.12979356-0.6446844745 .192499 \mathrm{e}-01$ factor(skip)2203-0.195064716 0.12387778 -1.574654639 1.155890e-01 factor(skip)2212 $0.0433651600 .160826960 .2696386257 .874828 \mathrm{e}-01$ factor(skip)2220 $0.1705528460 .486218440 .3507741227 .258169 \mathrm{e}-01$ factor(skip)2236-0.240310005 0.20254007-1.186481294 2.356576e-01 factor(skip)2248-0.468803577 0.35307606 -1.327769382 1.844965e-01 factor(skip)2265-0.081118608 $0.13404819-0.6051451275 .451923 \mathrm{e}-01$ factor(skip)2410 -0.192166371 $0.21987506-0.8739798343 .822970 \mathrm{e}-01$ factor(skip)2549-0.144974322 $0.14686104-0.9871530313 .237585 \mathrm{e}-01$ factor(skip)2550 $0.0275683340 .48384311 \quad 0.0569778379 .545719 \mathrm{e}-01$ factor(skip)2592 $0.2291146610 .35461605 \quad 0.6460921985 .183381 \mathrm{e}-01$ factor(skip)3033-1.789786537 $0.44139343-4.0548553825 .326475 \mathrm{e}-05$ factor(skip)3135-0.358279004 0.40653274 -0.881304184 3.783225e-01 factor(skip) $3156-1.2769434310 .36406219-3.5074870844 .683906 \mathrm{e}-04$ factor(skip) $3382-1.1742600490 .37396205-3.140051352$ 1.728735e-03 factor(skip) $3523-1.4154808450 .45907228-3.0833506962 .091640 \mathrm{e}-03$ factor(skip)3542-0.818668375 $0.38925897-2.1031458413 .565233 \mathrm{e}-02$ factor(skip) $3709-1.2127297190 .37572436-3.2277112211 .280216 \mathrm{e}-03$ factor(skip)934 -0.560484938 $0.10397540-5.3905533168 .388350 \mathrm{e}-08$ factor(skip)A $\quad 0.4788879000 .44197157 \quad 1.083526488$ 2.787836e-01 factor(skip)B $\quad 0.7474847330 .39901053 \quad 1.873345878$ 6.125365e-02 ---
Signif. codes: $0^{\text {'***' }} 0.001^{\text {'**' }} 0.01^{\text {'*' }} 0.05^{\text {'.' }} 0.1^{\text {' ' }} 1$
(Dispersion parameter for gaussian family taken to be 0.2216236 )
Null deviance: 1901.61 on 1335 degrees of freedom Residual deviance: 274.59 on 1239 degrees of freedom AIC: 1873.7

Number of Fisher Scoring iterations: 2
Analysis of Deviance Table
Model: gaussian, link: identity
Response: lafli

Terms added sequentially (first to last)

Df Deviance Resid. DfResid. Dev F $\quad \operatorname{Pr}(>F)$
ltogtimi $11353.661347559 .666153 .6477<2.2 \mathrm{e}-16^{* * *}$
factor(land) $4 \quad 68.12 \quad 1343 \quad 491.55 \quad 77.4139<2.2 \mathrm{e}-16^{* * *}$
factor(yy) $19 \quad 132.78 \quad 1324 \quad 358.76 \quad 31.7700<2.2 \mathrm{e}-16^{* * *}$
factor(mm) $9 \begin{array}{llllll} & 17.82 & 1315 & 340.94 & 9.0021 & 2.899 \mathrm{e}-13\end{array}{ }^{* * *}$
factor(skip) $64 \quad 65.75 \quad 1251 \quad 275.19 \quad 4.6701<2.2 \mathrm{e}-16^{* * *}$
---
Signif. codes: 0 '***' $0.0011^{\text {'**' }} 0.01^{\text {'*' }} 0.05^{\text {'.' }} 0.1^{\text {' ' }} 1$

Table 22.6.1 Deep pelagic $S$. mentella. Survey estimates for depth $>500 \mathrm{~m}$ from trawl samples taken in 2015. Areas C-F (Figure 22.6.2) were not surveyed.

|  | A | B | C | D | E | F |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Area (NM2) | 113450 | 87994 |  |  |  | TotaL |
| Mean length <br> $(\mathrm{cm})$ | 38.6 | 37.2 |  |  | 201444 |  |
| Mean weight $(\mathrm{g})$ | 673 | 668 |  |  | 68.3 |  |
| Biomass $(\mathrm{t})$ | 152775 | 64234 |  |  | 195694 |  |

Table 22.6.2. Results (biomass in ' 000 t ) for the international redfish surveys conducted since 1999 for deep pelagic S. mentella for each subarea (see Figure 22.6.2) and total. Areas C-F were not surveyed in 2015

| SUB-AREA |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | A | B | C | D | E | F | TOTAL |
| 1999 | 277 | 568 | 12 | 27 | 52 | 0 | 935 |
| 2001 | 497 | 316 | 28 | 79 | 64 | 18 | 1001 |
| 2003 | 476 | 142 | 20 | 13 | 27 | 0 | 678 |
| 2005 | 221 | 95 | 0 | 8 | 65 | 3 | 392 |
| 2007 | 276 | 166 | 1 | 5 | 62 | 11 | 522 |
| 2009 | 291 | 121 | 0 | 8 | 37 | 1 | 458 |
| 2011 | 342 | 112 | 0 | 1 | 18 | 0 | 474 |
| 2013 | 193 | 75 | 0 | 2 | 10 | 0 | 280 |
| 2015 | 153 | 43 | - | - | - | - | 196 |

### 22.13 Figures



Figure 22.2.1 Fishing areas and total catch of deep pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters 1992-2014. Data are from the Faroe Islands (1995-2013), Germany (2011-2014) Greenland (1999-2003 and 2009-2010), Iceland (1995-2014), and Norway (1995-2003 and 2010-2014). The catches in the legend are given as tones per square nautical mile. The blue box represents the proposed management unit.


Figure 22.2.1 (Cont.) Fishing areas and total catch of deep pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters 1992-2015. Data are from the Faroe Islands (1995-2015), Germany (2011-2015) Greenland (1999-2003 and 2009-2010), Iceland (1995-2015), and Norway (1995-2003 and 2010-2014). The catches in the legend are given as tones per square nautical mile. The blue box represents the proposed management unit.


Figure 22.2.2 Landings of deep pelagic S. mentella (Working Group estimates, see Table 21.2.1).


Figure 22.2.3 Trends in standardized cpue of the deep pelagic S. mentella fishery in the Irminger Sea and adjacent waters, based on logbook data from Faroe Islands, Iceland, Germany, Greenland and Norway.


Figure 22.2.4 Residuals from the GLM model used to standardize cpue, based on logbook data from Faroe Islands, Iceland, Greenland and Norway.


Figure 22.3.1 Length distribution from Icelandic landings of deep pelagic S. mentella.


Figure 22.6.1 Sub-areas A-F used on international surveys for redfish in the Irminger Sea and adjacent waters, and divisions for biological data (Northeast, Southwest and Southeast; boundaries marked by broken lines).


Figure 22.6.2. Redfish trawl estimates deeper than 500 m (type 3 trawls). $\mathrm{s}_{\mathrm{A}}$ values calculated by the trawl method (see WGRS Report, 2013) during the joint international redfish survey in June/July 2013.


Figure 22.6.3 Length distribution of redfish 1999-2015 in the trawls, by geographical areas (see Figure 22.6.1) and total, from fish caught deeper than 500 m .

## 23 Greenlandic slope Sebastes mentella in 14.b

### 23.1 Stock description and management units

See chapter 18 for description of the stock structure of $S$. mentella in the Irminger Sea and adjacent waters. ICES has advised separately for S. mentella found demersal in ICES 14.6 since 2011, and will do so until all available information on stock origin in this area is analysed and a new procedure is agreed upon.

### 23.2 Scientific data

Indices were available from three surveys in 14.b. A German survey directed towards cod in Greenlandic waters (0-400 meters) (Fock et al. 2015), the Greenland deep-water survey (400-1500 meters) targeting Greenland halibut (Christensen and Hedeholm 2016) and the Greenland shrimp and fish survey in shallow water ( $0-600$ meters), which has been conducted since 2008 (Hedeholm and Christensen. 2016). The German survey on the slope in 14. b has since 1982 been covering the slopes in East Greenland waters. Cod is the target species in this survey and it operates at depths of 400 meters and shallower. The survey was re-stratified in 2009 (see Stock Annex). From 1993-1998 a large number of Sebastes spp. smaller than 17 cm was found in the German survey (Figure 23.2.1). This coincided with a large increase in the amount of $17-30 \mathrm{~cm}$ large $S$. mentella from 1995-1998. From 1998 to 2003 the total biomass increased as a result of many small fish ( $<17 \mathrm{~cm}$ ) in the survey, followed by a few years of high biomass estimates for S. mentella from 2003-2009. This increase occurred in one particular stratum only, i.e. stratum 8.2 From 2009 onward, a declining trend in both biomass and abundance was observed, with 2015 representing the lowest biomass for the last 20 years (Figure 23.2.1). In the same period, the amount of small fish ( $17-30 \mathrm{~cm}$ ) has steadily declined causing an increase in the amount of larger fish (Figure 23.2.1) until the overall biomass declines in 2010 and 2011. The depletion of the small size group has led to a progressive decline in the juvenile biomass index to a current low level, and no new recruits have been seen in the survey since 2012. This pattern is also reflected in the abundance estimates (Figure 23.2.1). The modal size of the adult fish has increased from 25 cm in 2001 to around 37 cm in 2010, but declined slightly in 2011. The distribution has become flat with no clearly defined mode in 2013-2015 (Figure 23.2.2).

The Greenland deep-water survey has since 1998, except in 2001, surveyed the slopes of east Greenland from 400 to 1500 meters with the majority of stations deeper than 600 meters targeting Greenland halibut. The biomass indices in the Greenland deep-water survey peaked in 2012, but has decreased since then (Figure 23.2.3). The overall length distribution from the entire area in 2013 and 2014 shows a mode around 31 cm . In 2015, the mode increased slightly (Figure 23.2.4).

The Greenland shrimp and fish survey in shallow water in East Greenland started in 2007, and surveys the East Greenland shelf and shelf edge at depths between 0-600 meters. However, 2007 was mostly exploratory and is not reported. In general, survey estimates of schooling fish are associated with large uncertainties due to their patchy distribution. This, in conjunction with the relatively short time-series, makes overall conclusions regarding stock trends based solely on this survey tentative although it is probably the survey with the best coverage of redfish distribution. The 2015 biomass estimate for $S$. mentella was the lowest observed in the time-series (Figure 23.2.5). The German survey shows very similar trends both with regards to adult fish and juveniles. The juveniles are at the lowest level in the 30-year time-series, and the adult biomass index is declining and is at the lowest level for the last 20 years. Both survey length
distributions showed no clear mode, but a rather flat distribution (Figure 23.2.6). The German survey and the Greenland shrimp and fish shallow water survey both show overall declines in the $S$. mentella biomass

### 23.3 Information from the fishing industry

### 23.3.1 Landings

From the Greenland and German surveys we know that the demersal redfish found on the Greenland slope is a mixture of S. norvegicus and S. mentella. Based on the surveys and 11 samples from the commercial fishery the 8539 tonnes of demersal redfish caught in ICES $14 . b$, was estimated to be $70 \%$ S. mentella ( 5977 ) and $30 \%$ S. norvegicus ( 2562 ). Together with the proportion in 2014, this is the smallest proportion of S. mentella observed, which can be interpreted as $S$. norvegicus is becoming more abundant in the area. Prior to 1974, all catches were reported as S. norvegicus and the split was determined by working groups on a yearly bases.

Total annual landings of demersal S. mentella from Divisions $14 . \mathrm{b}$ since 1974 are presented in Table 23.3.1.1. From 1976-1994 annual landings were at a relatively high level with landings ranging between 2000 tons to 20000 tons with a very high peak at nearly 60000 t in 1976. However, this fishery was ended abruptly in 1995, due to large amounts of very small redfish in the catches. From 1998-2002 the landings ranged from 1000 to 2000 tonnes and from 2003 to 2008 landings remained at lower levels ( $<500$ tonnes). In 2009, an exploratory fishery landed 895 tonnes of S. mentella. This was a large increase compared to 2008 and for the first time in ten years the fishery was limited by a TAC. In 2010, a quota on 5000 tonnes demersal redfish was initially given and of these, 400 tonnes were allocated to the Norwegian fleet. After this amount was fished, an extraordinary research quota of 1000 tonnes was given to a Greenlandic vessel. Since 2010, the catches have been around 8300 tonnes (S. mentella and S. norvegicus combined) and in 2015 catches were 8539 tonnes (Figure 23.3.1.1). Since 2011 the TAC has been 8500 tonnes. In 2010, there was no jurisdiction that clearly delimited the pelagic stocks from the redfish found on the shelf. A few vessels benefitted from this by fishing their pelagic quota on the shelf ( 2179 tonnes) making catches on the shelf exceed the TAC. This led to the introduction of a "redfish line" that separates the demersal slope stock from the pelagic stocks (see stock annex).

### 23.3.2 cpue and bycatch cpue

A redfish bycatch cpue was introduced at the redfish 2012 benchmark (WKRED). This is based on catches from the Greenland halibut directed fishery (Christensen and Hedeholm 2016), which covers redfish distribution better than data from the redfish directed fishery and covers a longer period (1999-2015). The cpue has very low values in the initial two years of the time-series, but following an increase in 2001, values have remained at the same level until 2006 after which a decline followed. From 2010 to 2012, the cpue increased, followed by a decline in 2013-2015 (Figure 23.3.2.1).

The decline in the index follow the decline in biomass index seen in the shallow water surveys (German and Greenland). The Greenland halibut fishery is not as spatially restricted as the redfish fishery, thus it will not be as sensitive to local changes. Based on the cpue a decline in stock size seem to be present.

The cpue from the redfish directed fishery showed a drastic decline from $2010(3.7 \mathrm{t} / \mathrm{h})$ to 2015 ( $1.1 \mathrm{t} / \mathrm{h}$ ) (Figure 23.3.2.2). The fishery takes place in a geographically limited
area between $63.5^{\circ} \mathrm{N}$ and $65^{\circ} \mathrm{N}$, where approximately $90 \%$ of the catches are taken. Accordingly, the cpue series can only be used as an index on local stock development. Both the Greenland shallow water survey ( $0-600 \mathrm{~m}$ ) and the German survey ( $0-400 \mathrm{~m}$ ) show that the main fishing area coincides with the area of highest overall abundance. Hence, the cpue decline indicates a severe local stock depletion, which is also reflected in the overall stock trend.

### 23.3.3 Fisheries and fleets

The fishery for $S$. mentella on the slopes in $14 . \mathrm{b}$ is mainly conducted with bottom trawl, only about $1 \%$ were caught with longlines. The area where S. mentella is caught, is closely related to the area where fishery for Greenland halibut and cod takes place (Figure 23.3.3.1). The majority of the catches are taken at depths from $300 \mathrm{~m}-400 \mathrm{~m}$.
The directed fishery was stopped in 1995, but in 1998 Germany restarted a directed fishery for redfish with annual landings of approximately 1000 tonnes in 1998-2001 increasing to 2100 tonnes in 2002 (Bernreuther et al. 2013). Samples taken from the German fleet indicated that substantial quantities of the redfish caught, especially in 2002, were juveniles, i.e. fish less than 30 cm . There was very little demersal redfish fishery in 14.b in 2003-2004 (less than 500 tonnes). This continued in 2005-2008 and most $S$. mentella were caught as bycatch in the Greenland halibut fishery.

After the German fleet stopped fishing in 2002 the majority of the catches have been taken by the British, Faroese, Norwegian and Greenland fleet. The British fishery took place from 2001-2005 and since 2006 only Greenland, Faroese Islands, Norway and Germany have had any significant catches. In 2015, also Russia had a notable catch in the area (Table 23.3.3.2).

In 2009, three Greenland vessels started a fishery targeting demersal redfish. Each was given an explorative quota of 250 tonnes. This fishery was very successful and led to an increased fishery in 2010 (seven boats), 2011 ( 15 boats) and 2012 ( 21 boats). However, in $2012,95 \%$ of the catch was taken by six vessels and $97 \%$ by five vessels in 2013.

On the steep slopes very little horizontal distance separates the distribution of cod, redfish and Greenland halibut (Figure 23.3.3.2). The part of the fleet with both quotas for redfish and Greenland halibut takes advantage of this by shifting between very short hauls targeting redfish and long hauls directed to Greenland halibut. Thereby avoiding time where the vessel is not fishing due to processing of the catch.

### 23.3.4 Bycatch/discard in the shrimp fishery

To minimize bycatch of fish species in the fishery for shrimp the trawls have since 2002 been equipped with grid separators (G.H. 2001). However, the 22 mm spacing between the bars in the separator allows small fish to enter the codend. In a study on the amount of bycatch in the shrimp fishery the mean length of the redfish that entered the codend was $13-14 \mathrm{~cm}$. The same study also documented that redfish by weight accounted for less than $1 \%$ of the amount of shrimp that were caught (Sünksen 2007). Coincident with the introduction of these separator grids the amount of juvenile redfish caught by the shrimp fishery dropped from annual 100-200 tons to a lower level near 100 tonnes. Since 2006 limited shrimp fishery has taken place in ICES 14.b and the current level of bycatch must be considered negligible, and have for the last two years been 0 (Table 23.3.4.1). Since 1999, the fishery has started in April-May due to poor winter conditions such as ice and wind that prevents fishing. Only in 2000 and 2002, the fishery started already in February (Table 23.3.4.2). Since 2010, the fishery has been starting already in

January. The depth distribution of cod and redfish overlaps (Figure 23.3.3.2) and therefore the fishery for redfish led to a bycatch of cod on 96 tonnes in 2013. The vessels are allowed a $10 \%$ bycatch of cod.

### 23.3.5 Sampling from the commercial fishery

In 2015, the catch length distribution was estimated from 702 redfish and separated into S. mentella ( $\mathrm{N}=560$ ) and S. norvegicus $(\mathrm{N}=142)$ (Figure 23.3.5.1). The distribution showed a clear mode around 36 cm which is an increase compared to 2014. All samples were analysed by the Greenland Institute of Natural Resources, and it was found that S. mentella constituted $70 \%$ of the total sample weight. For S. norvegicus the mode was around 40 cm in 2015.

### 23.4 Methods

No analytical assessment was conducted.

### 23.5 Reference points (Benchmark, WKRED)

There are no biological reference points defined for this stock. However, part of the benchmark in 2012 (WKRED) was to evaluate the possible use of a stock production model in generating a quantitative advice for this stock. Under certain assumptions and for various intrinsic growth rates (r), current sustainable yields (and MSY) were calculated using the German survey and landings as input data. Across the range of r's, results seemed robust (CV range: 0.03-0.17), and the current sustainable yield was estimated at approximately 3.5 Kt . However, this procedure was criticized at the benchmark due to lack of coverage of redfish distribution in the survey and questionable landings, and it is stated in the benchmark report that: "The panel does not suggest that the Schaefer model approach used here is to be final; to the contrary it is offered as a first step (from which interim management advice might be formulated)". As there are doubts on stock structure, species determination (and hence catch data accuracy), migration and the quality of the surveys used as basis for the model approach, the applicability of the proposed reference points from WKRED is questionable. Indeed, the use of a stock production model on an aggregation of fish that is not clearly defined as a stock is questionable.

### 23.6 State of the stock

The German survey and the Greenland shrimp and fish shallow water survey both show overall declines in the S. mentella biomass since 2010, and both show a complete absence of small fish since $2013(<18 \mathrm{~cm})$. The adult stock decline is caused by a large decline in a small area which coincides with the main fishing area. The directed fishery cpue for this area has declined from $3.7 \mathrm{t} / \mathrm{h}(2010)$ to $1.1 \mathrm{t} / \mathrm{h}$ suggesting a large local decline. Changes in length distributions in both surveys also suggests that no new cohorts are present on the slope and that the adult biomass decline is caused by the gradual decline of a single/few cohorts. Especially the complete absence of juveniles is cause for concern.

The biomass estimate declines and the concentrated fishery could point to a fishery induced decline. However, the declines are of a magnitude that seems beyond what a limited number of years' catches can cause. Hence, surveys may either overestimate the biomass in especially Q3, not survey the entire area of distribution or S. mentella is disappearing due to migration. Survey overestimation may result from the large aggregations of redfish in Q3, which may cause two different survey scenarios, a low-
density and high-density situation. If large redfish aggregations change the catchability, the assumptions of linearity between catch and abundance are rendered invalid high fish concentration may simply reduce the trawl escape potential. Such a situation would produce disproportionally high catches and subsequently biomass estimates in high density areas such as Q3. Hence, the decline may be a synergetic effect of a reduced biomass caused by the local fishery, and the reduced catchability inferred from the less dense fish aggregations following some years of intense fishing. This is further complicated by the lack of knowledge of the stocks connection to the pelagic (deep and shallow) and Icelandic slope stocks and the degree of migration. Based on this, care must be taken when evaluating stock status, but nevertheless, the consistency in both the German and shallow Greenlandic surveys suggests that the biomass has decreased, especially in area Q3. The magnitude of the decline is probably not attributable to the fishery alone. Also, the apparent lack of juveniles in all the East Greenland area means that no new fish will grow into the fishable part of the stock for at least 6-8 years, and there is reason for concern.

The advice is based on the Data Limited Stock approach (DLS) including biomass indices from the Greenland shallow water survey in the most recent 5 years combined with the recent advice. Due to the dynamic of the stock and decrease in all stock indicators the indicator ratio ( $50 \%$ ) was used and the uncertainty cap not applied. According to the guidelines the precautionary buffer was not applied. The advice for 2017 is 1120 tonnes.

### 23.7 Management considerations

Sebastes mentella is a slow growing, late maturing deep-sea species and is therefore considered vulnerable to overexploitation and advice has to be conservative. The fact that the fishery is targeting a localized aggregation of fish is cause for concern as is the absence of juveniles in the area. Given the biology of the species and the uncertainty in the biomass trend, any advice should consider this a hot spot fishery as it is potentially detrimental to this local and potentially important aggregation of larger fish. The fishery should still be at a low level involving few vessels. This should be maintained until the effect of the fishery can be clarified, especially with the recent declines in biomass estimates and the fishery should preferably cover a larger area.

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### 23.9 Tables

Table 23.3.1.1 Nominal landings (tonnes) of demersal S. mentella 1974-2013 ICES division 14.b.

| Demersal S.mentella |  |
| :---: | :---: |
| 1974 | 0 |
| 1975 | 4400 |
| 1976 | 59700 |
| 1977 | 0 |
| 1978 | 5403 |
| 1979 | 5131 |
| 1980 | 10406 |
| 1981 | 19391 |
| 1982 | 12140 |
| 1983 | 15207 |
| 1984 | 9126 |
| 1985 | 9376 |
| 1986 | 12138 |
| 1987 | 6407 |
| 1988 | 6065 |
| 1989 | 2284 |
| 1990 | 6097 |
| 1991 | 7057 |
| 1992 | 7022 |
| 1993 | 14828 |
| 1994 | 19305 |
| 1995 | 819 |
| 1996 | 730 |
| 1997 | 199 |
| 1998 | 1376 |
| 1999 | 853 |
| 2000 | 982 |
| 2001 | 901 |
| 2002 | 2109 |
| 2003 | 446 |
| 2004 | 482 |
| 2005 | 267 |
| 2006 | 202 |
| 2007 | 226 |
| 2008 | 92 |
| 2009 | 895 |
| 2010 | 6613 |
| 2011 | 6705 |
| 2012 | 6572 |
| 2013 | 6597 |
| 2014 | 4608 |
| 2015 | 5977 |

Table 23.3.3.2 Landings (tonnes) of demersal redfish caught in ICES 14.b by nation. By far the largest proportion were probably $S$. mentella but none of these amounts were converted by the mentella/norvegicus ratio ( $70 \%$ S. mentella) found by the two surveys covering the area.

| YEAR | DEU | ESP | EU | FRO | GBR | GRL | ISL | NOR | POL | RUS | UNK |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: | ---: |
| 1999 |  |  |  |  |  |  |  |  |  | SUM |  |
| 2000 | 884 |  | 11 |  |  | 19 | 65 | 853 | 853 |  |  |
| 2001 | 782 |  |  |  | 11 | 9 |  | 99 |  | 3 | 982 |
| 2002 | 1703 |  |  | 48 | 16 | 246 | 29 | 32 |  | 36 | 901 |
| 2003 | 3 | 2 | 2 | 20 | 155 | 232 |  | 32 |  |  | 2109 |
| 2004 | 5 | 1 | 79 | 12 | 221 | 93 |  | 68 | 3 |  | 446 |
| 2005 | 2 |  | 4 | 38 | 96 | 72 |  | 56 |  |  | 482 |
| 2006 | 1 |  |  |  |  | 152 |  | 48 |  |  | 267 |
| 2007 | 7 |  | 15 | 138 |  | 35 |  | 30 |  |  | 202 |
| 2008 | 1 |  | 8 | 50 | 5 | 5 |  | 23 |  |  | 226 |
| 2009 |  |  |  | 203 |  | 822 |  | 93 |  |  | 92 |
| 2010 | 10 |  | 12 | 381 |  | 5672 |  | 2190 |  | 1 | 1118 |
| 2011 | 1262 |  | 26 | 2 |  | 6757 |  | 334 |  | 1 | 8266 |
| 2012 | 1810 |  | 5 | 32 |  | 5964 | 1 | 403 |  | 1 | 8381 |
| 2013 | 1957 |  |  | 32 | 30 | 5863 |  | 356 |  | 8 | 8216 |
| 2014 | 1973 |  | 0.2 | 13 |  | 4611 | 98 | 613 |  | 5 | 8246 |
| 2015 | 1987 |  | 74 |  | 4979 | 208 | 822 |  | 469 | 7314 |  |
| Sum | 12387 | 3 | 162 | 1043 | 534 | 35531 | 336 | 5264 | 3 | 521 | 856 |

Table 23.3.4.1 Discarded bycatch (tonnes) of Sebastes sp. from the shrimp fishery in ICES 14.b.

| YEAR | JAN | FEB | MAR | APR | MAY | JUN | JUL | AUG | SEP | OCT | NOV | DEC | SUM |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 6 | 16 | 17 | 5 | 1 | 13 | 2 | 48 | 22 | 30 | 40 | 33 | 234 |
| 2000 | 10 | 3 | 31 | 17 | 15 | 4 | 21 | 78 | 28 | 18 | 9 | 6 | 239 |
| 2001 | 7 | 9 | 10 | 16 | 9 | 11 | 4 | 5 | 3 | 3 | 28 | 6 | 111 |
| 2002 | 3 | 11 | 9 | 6 | 1 | 0 | 0 | 5 | 4 | 8 | 3 | 5 | 55 |
| 2003 | 5 | 6 | 8 | 5 | 5 | 8 | 8 | 15 | 2 | 10 | 12 | 4 | 88 |
| 2004 | 7 | 10 | 17 | 13 | 4 | 2 | 27 | 20 | 7 | 2 | 9 | 0 | 118 |
| 2005 | 7 | 14 | 16 | 8 | 7 | 5 | 6 | 21 | 14 | 4 | 5 | 20 | 126 |
| 2006 | 6 | 2 | 4 | 1 | 3 | 5 | 2 | 4 | 4 | 0 | 0 | 4 | 35 |
| 2007 | 7 | 3 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 |
| 2008 | 0 | 2 | 2 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 7 |
| 2009 | 1 | 2 | 11 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 16 |
| 2010 | 1 | 2 | 2 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 2 | 10 |
| 2011 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2012 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2013 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sum | 60 | 81 | 131 | 75 | 48 | 49 | 71 | 196 | 84 | 75 | 106 | 81 | 1056 |

Table 23.3.4.2 Landings (tonnes) of demersal redfish caught in ICES 14.b. by month. By far the largest proportion were probably S. mentella but none of these amounts were converted by the mentella/norvegicus ratio $(80 \%$ S. mentella) found by the two surveys covering the area

| Year | JAN | FEB | MAR | APR | MAY | Jun | JuL | AUG | SEP | OCt | NOV | DEC | SUM |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 |  | 10 |  | 108 |  | 4 | 42 | 10 | 15 | 34 | 481 | 149 | 853 |
| 2000 | 18 | 238 | 286 | 260 | 10 | 4 | 79 | 72 | 13 | 0 | 3 |  | 982 |
| 2001 |  |  | 1 |  |  |  | 108 | 2 |  | 184 | 369 | 236 | 901 |
| 2002 |  | 183 | 445 | 354 | 390 | 50 | 472 | 35 | 44 | 59 | 77 |  | 2109 |
| 2003 |  |  | 9 | 4 | 26 | 27 | 135 | 195 | 20 | 16 | 12 |  | 446 |
| 2004 |  |  |  | 35 | 41 | 63 | 75 | 48 | 64 | 96 | 25 | 35 | 482 |
| 2005 |  |  | 1 | 15 | 66 | 24 | 80 | 29 | 13 | 18 | 19 |  | 267 |
| 2006 |  | 3 | 7 | 50 | 14 | 39 | 20 | 61 | 2 | 1 | 1 | 2 | 202 |
| 2007 | 6 | 13 | 8 | 8 | 14 | 42 | 4 | 106 | 16 | 7 | 1 | 1 | 226 |
| 2008 | 4 | 3 | 1 | 6 | 12 | 11 | 31 | 12 | 10 | 2 |  |  | 92 |
| 2009 |  |  |  | 1 | 84 | 346 | 148 | 105 | 128 |  | 288 | 17 | 1118 |
| 2010 | 799 | 786 | 708 | 1058 | 2149 | 2100 | 108 | 134 | 88 | 301 | 36 |  | 8266 |
| 2011 | 419 | 1396 | 1661 | 1017 | 268 | 250 | 236 | 598 | 255 | 583 | 1223 | 475 | 8381 |
| 2012 | 899 | 2197 | 628 | 852 | 577 | 699 | 966 | 143 | 44 | 23 | 474 | 712 | 8215 |
| 2013 |  |  | 709 | 1290 | 925 | 1423 | 1218 | 1086 | 723 | 227 | 119 | 527 | 8246 |
| 2014 | 10 | 421 | 206 | 1210 | 1187 | 1709 | 231 | 401 | 376 | 448 | 632 | 479 | 7314 |
| 2015 | 543 | 786 | 1016 | 451 | 507 | 1611 | 1160 | 1024 | 504 | 393 | 74 | 467 | 8539 |
| Sum | 2698 | 6036 | 5686 | 6719 | 6270 | 8402 | 5113 | 4061 | 2315 | 2392 | 3834 | 3100 | 56640 |

### 23.10Figures



Figure 23.2.1. Indices from the German East Greenland survey of S. mentella larger than 17 cm . Biomass (a), abundance (b), and biomass split on length (c). On figure (c) the grey bars represent the biomass of $S$. mentella larger than 30 cm and the light bars biomass in fish from 17-30 cm .


Figure 23.2.2. Length distributions from the German East Greenland survey 1985-2015.


Figure 23.2.3. Biomass of S. mentella and Sebastes spp. derived from the deep Greenland survey. Bars indicate 2SE of the biomass of S. mentella including Sebastes spp. No survey in 2001. In 2004, 2005 and 2007 a large proportion of the redfish were not determined to species and only reported as "Sebastes spp". It is most likely that the majority of these fish were S. mentella.


Figure 23.2.4. Overall length distribution of Sebastes mentella (number per $\mathbf{k m}^{\mathbf{2}}$ ) from the deep Greenland survey.


Figure 23.2.5: Biomass ( $\mathrm{kg}^{*} 10^{6}, \mathrm{Kt}$ ) $( \pm \mathrm{CV} \%)$ indices for S. mentella (top) and Sebastes sp. $(<18 \mathrm{~cm})$ (bottom) off East Greenland in 2008-2015 from the Greenlandic shallow water survey. All surveyed areas (Q1-Q6) are combined.


Figure 23.2.6. Overall length distributions for juvenile redfish S. mentella (left) and Sebastes spp. $<18 \mathrm{~cm}$ (right) (note the change in scale from 2013) from the Greenlandic shallow water survey. All surveyed areas combined (Q1-Q6).


Figure 23.3.1.1 Landings of S. mentella in subarea 14.b. Landings of "redfish" have been split based on estimates from survey and commercial catches.


Figure 23.3.2.1 Standardized redfish bycatch cpue in the directed fishery for Greenland halibut in ICES 14.b as a function of year. cpue was estimated from the GLM model: Incpue=year+ICES Subdivision+depth. Bars represent standard error. Only hauls made below 1000 m were used in the analyses.


Figure 23.3.2.2 Standardized redfish cpue in the redfish directed fishery ICES 14.b as a function of year. cpue was estimated from the GLM model: Incpue=year+ICES Subdivision+depth. Dashed lines represent standard error.


Figure 23.3.3.1 Distribution of catches of demersal redfish in 2015 in ICES 14.b.


Figure 23.3.3.2. Lines represent the share of the total commercial catch caught at a given depth from 1999-2011 in G. morhua, demersal redfish (mixed S. mentella and S. norvegicus) and R. hippoglossoides.


Figure 23.3.5.1: Length distribution of 702 redfish analysed by the Greenland Institute of Natural Resources separated into S. mentella ( $\mathrm{N}=559$ ) and S. norvegicus $(\mathrm{N}=143$ ).

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27 April - 05 May 2016

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## Annex 2: ToRs for the Next Meeting

The North-Western Working Group (NWWG), chaired by Rasmus Hedeholm, Greenland, will meet at ICES Headquarters, 26 April - 3 May, 2017 to:
a) Address generic ToRs for Regional and Species Working Groups.

The assessments will be carried out on the basis of the stock annex. The assessments must be available for audit on the first day of the meeting. Material and data relevant to the meeting must be available to the group no later than 5 April 2017 according to the Data Call 2017.

For capelin in Iceland-East Greenland-Jan Mayen area, NWWG will agree any changes to the WG type report and the draft advice no later than 10 May 2017.

The assessments will be carried out on the basis of the stock annex in National Laboratories, prior to the meeting. For capelin in Iceland-East Greenland-Jan Mayen area, Iceland will provide a WG type report and a draft advice sheet on 3 May. NWWG will agree any changes to the WG type report and the Advice sheet no later than 10 May. An ADG will work by correspondence XX May. The WEBEX will be XX May, and the Advice Release date XX May.
Other material and data relevant to the meeting must be available to the group no later than 14 days prior to the starting date. NWWG will report by XX May 2017 for the attention of ACOM. For capelin in Iceland-East Greenland-Jan Mayen area NWWG will report by XX February 2018 for the attention of ACOM.

## Annex 3: List of Stock Annexes

The table below provides an overview of the NWWG Stock Annexes. Stock Annexes for other stocks are available on the ICES website Library under the Publication Type "Stock Annexes". Use the search facility to find a particular Stock Annex, refining your search in the left-hand column to include the year, ecoregion, species, and acronym of the relevant ICES expert group.

| Stock ID | Stock name | Last updated | Link |
| :---: | :---: | :---: | :---: |
| cap-icel_SA | Capelin in the Iceland-East Greenland-Jan Mayen area) | January 2015 | cap-icel_SA.pdf |
| cod-farb_SA | Cod (Gadus morhua) in Subdivision 5.b2 (Faroe Bank) | April 2013 | cod-farb_SA.pdf |
| cod-farp_SA | Cod (Gadus morhua) in Subdivision 5.b1 (Faroe Plateau) | May 2016 | cod-farp_SA.pdf |
| cod-iceg_SA | Icelandic cod | January 2015 | cod-iceg_SA.pdf |
| cod-ingr_SA | Cod in inshore waters of NAFO Subdivision 1A-1F (Greenland cod) | January 2015 | cod-ingr_SA.pdf |
| cod-segr_SA | Offshore cod in South (NAFO Subdivision 1F) and East Greenland | May 2016 | cod-segr_SA.pdf |
| cod-wgr_SA | Offshore cod in West Greenland (NAFO Subdivision 1A-1E | May 2016 | cod-wgr_SA.pdf |
| ghl-grn_SA | Greenland halibut (Reinhardtius hippoglossoides) in Subareas 5, 6, 12, and 14 (Iceland and Faroes grounds, West of Scotland, North of Azores, East of Greenland) | December 2013 | ghl-grn_SA.pdf |
| had-faro_SA | Haddock (Melanoggrammus aeglefinus) in Division 5.b (Faroes grounds) | April 2014 | had-faro SA.pdf |
| had-iceg_SA | Haddock (Melanogrammus aeglefinus) in Division 5.a (Iceland) | February 2013 | had-iceg_SA.pdf |
| her-vasu_SA | Herring (Melanogrammus aeglefinus) in Division 5.a (Iceland) | February 2013 | her-vasu SA.pdf |
| sai-faro_SA | Saithe (Pollachius virens) in Division 5.b (Faroes grounds) | May 2016 | sai-faro_SA.pdf |
| sai-icel_SA | Saithe in Division Va (Icelandic waters) | May 2013 | sai-icel SA.pdf |
| smn-con_SA | Icelandic slope beaked redfish (Sebastes mentel/a) in Divisions 5.a and 14.b | May 2012 | smn-con_SA.pdf |
| smn-dp_SA | Deep Pelagic beaked redfish (Sebastes mentella) in ICES | May 2012 | smn-dp_SA.pdf |


| smn-grl_SA | Beaked redfish (Sebastes mentella) in Division 14.b (Demersal) (Southeast Greenland) | May 2016 | smn-grI_SA.pdf |
| :--- | :--- | :--- | :--- |
| smn-sp_SA | Shallow Pelagic Beaked Redfish (Sebastes mentella) | May 2012 |  |
| smn-5614_SA | Golden redfish in Subareas 5 and 14 (Iceland, East Greenland and Faroe Islands) | February 2014 | smn-sp SA.pdf |

# Audit of Cod (Gadus morhua) in NAFO Subarea 1, inshore (West Greenland cod) (cod-ingr). 

Date: 4 May 2016
Auditor: Petur Steingrund

## General

The report is well done although the language could be improved. The tables and figures are easy to understand and there is a considerable work done to highlight the different year classes and their spatial distribution. The basis for the advice has varied a lot during the years and is confusing because different ICES bodies have had different opinion. ICES should try to avoid this in the future.

## For single-stock summary sheet advice:

There is no analytical assessment done for this stock. A catch curve analysis is performed, which gives information about total mortality (very high). An age-based assessment was performed a few years ago, but it was discontinued.

1 ) Assessment type: Catch curve analysis
2 ) Assessment: Total mortality (very high)
3 ) Forecast: Presented. A linear regression of survey indices and subsequent catch (under the assumption that the catch was a relative indication of stock size). This forecast model was suggested by a benchmark to be used as a basis for the advice, but the model was rejected by the Advice Drafting Group.
4 ) Assessment model: No assessment model.
5 ) Data aresues: No particular data aresues, although a reliable survey index of the fishable stock is missing (there is just a reliable survey index of recruitment).

6 ) Consistency: No assessment model, so this is not relevant.
7 ) Stock status: Unknown.
8 ) Management Plan: There is no management plan for this stock.

## General comments

This stock has quite good data (catch-at-age and survey-at-age data) that could be used in an age-based stock assessment model. There is no reliable commercial cpue. The language could be improved and the frequent use of numbers hampers the reading. The mixture of cod stocks in the coastal areas may be a problem for the evaluation of Z -values and the perception of the stock status.

## Technical comments

The advice procedure (how the advice is reached) is not present in the annex. In the annex there is a description of the model that was previously used as the basis for the advice (result of the benchmark), but later rejected by the Advice Drafting Group. The
annex needs to be updated, e.g. the terminal year seems to be 2013 or 2014 and the table under the biological reference points has the notion "Explain".

## Conclusions

Given the relatively good data (catch-at-age and survey-at-age) it is unfortunate that an age-disaggregated stock assessment model is not used for this stock. Although such a model may be associated with problems it may still be much better than the current procedure ("you see a scratch in the Ferrari and then you take the bike instead"). An age based stock assessment model is also an excellent way of checking the consistency in the data, e.g. how well the cohort data fit to the model and to which extent certain year classes migrate away.

## Audit of Cod in Division 5.a (Icelandic cod; cod-iceg)

Date: 10 May 2016
Auditer: Anja Retzel

## General

The stock has been assessed in close agreement with the stock annex.

## For single-stock summary sheet advice:

1 ) Assessment type: update
2 ) Assessment: analytical
3 ) Forecast: presented
4 ) Assessment model: statistical catch-at-age (ADCAM) tuned with two (spring and fall) surveys.
5 ) Data aresues: All data are available as described in the Stock Annex.
6 ) Consistency: SPALY assessment was consistent with last year's assessment.
7 ) Stock status: Blim is 125 kt , MSY Btrigger is 220 kt and SSB in 2016 is estimated at 469 kt . Reference biomass, (B4+) is estimated at 1243 kt in 2016. $\mathrm{B}_{\mathrm{pa}}$ is 160 kt , $\mathrm{F}_{\text {lim }}$ is 0.74 and $\mathrm{F}_{\mathrm{pa}}$ is 0.69 .

8 ) Man. Plan: Because SSB> Btrigger, the TAC2016/2017 is set as (TAC2015/2016 + $\left.0.2^{*} \mathrm{~B}_{\mathrm{B} 4+2216}\right) / 2$. In accordance with this plan, the proposed TAC for 2016/2017 is 244 kt . According to the advice sheet, ICES has evaluated the plan and concludes that it is in accordance with the precautionary approach and the ICES MSY framework.

## General comments

This was a well-documented, well ordered and considered section. It was easy to follow and interpret.

## Technical comments

None

## Conclusions

The assessment has been performed in as close proximity to the Stock annex and the results can be used as basis for advice.

## Audit of Faroe Bank Cod (cod-farb)

Date: 10 May 2016
Auditor: Anja Retzel

## General

## For single-stock summary sheet advice:

1) Assessment type: Update
2) Assessment: trends
3) Forecast: not presented
4) Assessment model: None
5) Data aresues: None, in relation to Stock Annex

6 ) Consistency: Same advice as in last year, the area closed for fishery.
7) Stock status: Poor status of the stock, the biomass indices below average and no signs of improvements
8) Man. Plan.: There is no management plan for this stock

## General comments

The message in the text is clearly put forward and the graphs illustrative. However, the structure and subheadings of the Assessment report is a bit different from other stocks, and should be improved and made more along the lines with others at some point (e.g. during benchmark assessment). For example, the section "Status of the stock" is normally few sentences but fills 1.5 pages out of total 3 pages for this stock. Most of the text there should be under different sections (e.g. "Landings","Information from Surveys", "Exploratory assessment" etc).

## Technical comments

I did not see any errors in the report and the assessment was done according to the Stock Annex.

## Conclusions

The assessment has been performed correctly

## Checklist for review process

## General aspects

- Has the EG answered those TORs relevant to providing advice? -Yes
- Is the assessment according to the stock annex description? -Yes
- Is general ecosystem information provided and is it used in the individual stock sections. -Not provided in the stock report, and therefore not used
- If a management plan has been agreed, has the plan been evaluated? -Not relevant


## For update assessments

- Have the data been used as specified in the stock annex? -Yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex? -Not relevant, it is surveys trends based assessment
- Is there any major reason to deviate from the standard procedure for this stock? -No
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice? -Yes, the update assessment gives a valid basis for advice.


## Audit of Capelin in the Iceland-East Greenland-Jan Mayen area (capicel)

Date: 25 May 2016
Auditor: Teunis Jansen

## General

The basis for the advice was in accordance with the stock annex and the surveys were well described. However:

- Stock annex was missing in the 2015 report. It is available from the WKICE 2015 report, so the NWWG ensure that it is included in the 2016 report.
- Information about predation and growth data used as prediction model input is lacking. Discuss the estimates in relation to previous years. How sensitive is the forecast and thereby advice to uncertainties in predation. Did migration speed and route differ from the historic information used for the predation model?
- Assumptions about growth are not described (see Stock annex section C. Assessment: data and method, paragraph with the header "Following acoustic survey in autumn (September-October)": "Bootstrap replicates of survey estimates of SSB are updated based on revised assumptions about growth, mortality, etc. Additional uncertainty is included due to variable mortality and fed into the predation model with same supporting data as when applied to winter survey results.")


## Audit of Faroe plateau cod (cod-farp)

Date: 11 May 2016
Auditor: Helle Torp Christensen

## General

## For single-stock summary sheet advice:

1) Assessment type: Update

2 ) Assessment: XSA with catch-at-age data and age-disaggregated indices, using catches in the model and in the forecast.

3 ) Forecast: The short-term prediction until year 2018 showed a slightly decreasing total-stock biomass to 25000 tonnes and a spawning-stock biomass to 19000 tonnes. Single species long-term forecasts for Faroe Plateau cod indicated Fmsy values lower than Fpa. The 2016 assessment was much in line with the 2015 assessment and forecast

4 ) Assessment model: Short term: Age structure and Long term: Yield and biomass per recruit over a range of F -values.

5 ) Data aresues: Landing data are considered accurate. There are no incentives to discard fish under the effort management system. The sampling of the landings is believed to be adequate.
6 ) Consistency: Consistent with last year
7 ) Stock status: The stock is assessed to be in a very poor state and is predicted to remain so for the next two years due to poor recruitment.

8 ) Management Plan: There is no management plan for this stock.

## General comments

ToRs relevant to provide advice is presented and the report is well structured and text is clear. The stock is in very poor state and a management plan for recovery is suggested.

## Technical comments

Assessment was conducted according to the Stock Annex.

## Conclusions

The assessment has been performed correctly and the results can be used as basis for advice.

## Audit of Cod (Gadus morhua) in NAFO Subarea 1A-1E (Offshore West Greenland) (cod-wgr).

Date: 4 May 2016
Auditor: Alexey Rolskiy

## General

A comprehensive review of scientific and fishery data. Lots of detailed figures and tables including information about tagging. Some text corrections are needed.

## For single-stock summary sheet advice:

There is no analytical assessment done for this stock.
Assessment type: Qualitative.
1 ) Assessment: Survey trends analysis.
2 ) Forecast: The surveys (GRL-GFS, Ger(GRL)-GFS-Q4) show only a minor increase in biomass in recent years, but it is assumed that survey uncertainty is high.
3 ) Assessment model: No assessment model.
4 ) Data aresues: All data are available as described in the Stock Annex.
5 ) Consistency: No assessment model, so this is not relevant.
6 ) Stock status: The stock is considered to be at a very low level compared to historic.
7 ) Management Plan: implemented for the offshore cod fishery in Greenland (2014-2016), based on the distinction between the inshore and two offshore stocks components.

## General comments

No stock assessment can be undertaken for this stock, due to the lack of significant rebuilding since the stock collapsed in the late 1960s. Surveys indicate some increase of juveniles in the last decade, but it is believed that this increase is partly driven by juveniles from other cod stocks using the area as nursery grounds. This circumstance adds uncertainty to the utility of the survey indices. Commercial cpue data are available. However, due to the limited fisheries in recent years they are of little use for stock assessment.

## Technical comments

None

## Conclusions

Given the very low level of this stock, no catches should be taken in 2017. Further work should be invested into the data sources used for the assessment after any signs of significant rebuilding of the stock.

## Audit of Icelandic slope Sebastes mentella (smn-con)

Date: 11 May 2016
Auditor: Helle Torp Christensen

## General

## For single-stock summary sheet advice:

1 ) Assessment type: Survey trend-based assessment
2 ) Assessment: Trends
3 ) Forecast: not presented
4 ) Assessment model: No analytical assessment. Survey indices from the annual survey since 2000 are used as basis for advice.

5 ) Data aresues: No particular data aresues. However, systematic age reading from the collected otolites could improve knowledge of the stock.
6 ) Consistency: Not applicable due to no assessment.
7 ) Stock status: State of the stock is on a low level. With the information at hand, current exploitation rates cannot be evaluated for the Icelandic slope S. mentella in Division 5.a. The fishable biomass index of Icelandic slope $S$. mentella from the Icelandic autumn survey shows that the biomass index for 2004-2013 has decreased to similar level as in 2003 when it was at lowest level, but increased again in 2014 and 2015. The German survey and the Greenland shrimp and fish shallow water survey both show no new recruits ( $>18 \mathrm{~cm}$ ) and no juveniles are present $(<18 \mathrm{~cm})$.

8 ) Management Plan: There is no management plan for the stock.

## General comments

The report addresses the ToRs of the WD in proper and thorough manners, as achievable with these data limited stock. It is the perception of the stock that it is on a general low level, however it is stable or slightly recovering.

## Technical comments

The report is in accordance with the stock annex.

## Conclusions

The assessment has been performed correctly and is a valid basis for advice.

## Audit of Shallow Pelagic Sebastes mentella (smn-sp-SA)

Date: 2 May 2016
Auditor: Gudmundur J. Óskarsson

## General

## For single-stock summary sheet advice:

1 ) Assessment type: update
2 ) Assessment: trends
3 ) Forecast: not presented
4 ) Assessment model: No analytical assessment, -the assessment is based on survey indices, catches, cpue and biological data

5 ) Data aresues: No new survey data available from 2015 because incomplete survey coverage (detailed in the report), thus the newest survey data derived from 2013. Consequently, lack of survey data for so long period is in contrast to the description in the stock annex.
6 ) Consistency: Not applicable due to no assessment or no new survey/cpue data presented in the report.
7 ) Stock status: The most recent survey estimate, in relation to the time-series, indicate a really poor status of the stock

8 ) Management Plan: No management plan exists for this stock

## General comments

The report addresses the ToRs of the WD in proper and thorough manners, as achievable with these data limited stock. Due to lack of new information, particularly survey data, the changes in the report from previous year are small and consequently the perception of the stock size has not changed for some years. Considering the apparently very poor stock status, the countries still fishing from the stock must have obligations to participate in scientific surveys to assess the stock size. This last sentence reflects my thoughts after having read the report.

## Technical comments

There were minor errors spotted in the report (kt instead of million $t$ ), that have been passed on to the stock coordinator. Otherwise, the content of the report is in accordance with the stock annex.

## Conclusions

The assessment has been performed correctly

## Audit of Greenland slope Sebastes mentella in 14.b. (smn-grl)

Date: 4 May 2016
Auditor: Kristján Kristinsson

## General

## For single-stock summary sheet advice:

1 ) Assessment type: update
2 ) Assessment: trend based assessment
3 ) Forecast: not presented
4 ) Assessment model: No analytical assessment. The assessment is based on survey biomass index from the Greenland shrimp and fish shallow water survey (GRL-GFS, $0-600 \mathrm{~m}$ ). Other surveys survey indicators are German survey (GER(GRL)-GFS-Q4, 0-400 m), and Greenland deep-water survey (GRL-DEEP, 400-1500 m). Catches, cpue and biological data.

5 ) Data aresues: Catch statistics do not separate Sebastes mentella from Sebastes norvegicus. The survey indices have large uncertainties, primarily due to the aggregating behaviour of the stock. Connection to other stocks of S. mentella in the area (shallow and deep pelagic stocks and the Icelandic slope stock remains unresolved.

6 ) Consistency: Not applicable due to no assessment.
7 ) Stock status: Biomass and abundance index for both adult and juvenile redfish have together with the cpue been declining since 2010. Biomass estimates from the Greenland shrimp and fish shallow water survey were in 2015 the lowest observed. Stock indicator from the other two surveys also show decline.

8 ) Management Plan: No management plan exists for this stock

## General comments

The report addresses the ToRs of the WD in proper and thorough manners, as achievable with these data limited stock. Landings have in 2011-2015 been well above recommended TAC which is of concern because of sharp decline in stock indicators. Advice has to be conservative.

## Technical comments

## Conclusions

The assessment has been performed correctly and as described in the Stock Annex

## Annex 5: List of Working Documents. (NWWG 2016)

Boje J. and Hedeholm R. 2016. The fishery for Greenland halibut in ICES Div. XIVb in 2015. ICES NWWG 2016 Working Document no. 01.
Boje J., Hvingel C. and Hedeholm R. 2016. An assessment of Greenland halibut (Reinhardtius hippoglossoides) off East Greenland, Iceland and the Faroe Islands. ICES NWWG 2016 Working Document no. 02.

Óskarsson G.J. 2016. Estimation on number-at-age of the catch of Icelandic summer-spawning herring in 2015/2016 fishing season and the development of Ichthyophonus hoferi infection in the stock. ICES NWWG 2016 Working Document no. 03.
Óskarsson G.J. 2016. Results of acoustic measurements of Icelandic summer-spawning herring in the winter 2015/2016. ICES NWWG 2016 Working Document no. 04.

Steingrund P. 2016. Greenland halibut CPUE for commercial trawlers operating on the slope on the Faroe Plateau 1991-2015. ICES NWWG 2016 Working Document no. 05.
Steingrund P. 2016. Greenland halibut CPUE for the research vessel operating on the slope on the Faroe Plateau in May-June 1995-2015. ICES NWWG 2016 Working Document no. 06.
Steingrund P. 2016. Survey biomass indices of Greenland halibut on the slopes of the Faroe Plateau 1983-2015. ICES NWWG 2016 Working Document no. 07.

Steingrund P. 2016. A combined biomass index of Greenland halibut on the slopes of the Faroe Plateau 1983-2015. ICES NWWG 2016 Working Document no. 08.
Christensen H.T. and Hedeholm R. 2016. The fishery for demersal Redfish (S.mentella) in ICES Div. XIVb in 2015. ICES NWWG 2016 Working Document no. 09.

Hedeholm R. and Christensen H.T. 2016. Greenland Shrimp and Fish Survey Results for Redfish in East Greenland Offshore Waters in 2015. ICES NWWG 2016 Working Document no. 10.

Christensen H.T. and Hedeholm R. 2016. Survey for Greenland halibut in ICES Division 14B, August - September 2015. ICES NWWG 2016 Working Document no. 11.
Steingrund P. and Reinert J. 2016. The biomass of Faroe haddock 1914-1956 estimated by catch per unit effort and stock assessment data. ICES NWWG 2016 Working Document no. 12.

Steingrund P. 2016. The biomass of Faroe saithe 1924-1960 estimated by catch per unit effort and stock assessment data. ICES NWWG 2016 Working Document no. 13.
Popov V. and Rolskiy A. 2016. Preliminary information on the results of Russian fishery and biological samples of pelagic redfish from the ICES subarea XII, XIV and NAFO Div. 1F in 2014. ICES NWWG 2016 Working Document no. 14.

Retzel A. 2016. Greenland commercial data for Atlantic cod in East Greenland offshore waters for 2015. ICES NWWG 2016 Working Document no. 15.
Retzel A. 2016. Greenland Shrimp and Fish survey results for Atlantic cod in ICES subarea XIVb (East Greenland) and NAFO subarea 1F (SouthWest Greenland) in 2015. ICES NWWG 2016 Working Document no. 16.

Retzel A. 2016. Greenland commercial data for Atlantic cod in Greenland inshore waters for 2015. ICES NWWG 2016 Working Document no. 17.

Retzel A. and Christensen H.T. 2016. West Greenland inshore survey results for Atlantic cod in 2015. ICES NWWG 2016 Working Document no. 18.

Retzel A. 2016. Greenland commercial data for Atlantic cod in West Greenland offshore waters for 2015. ICES NWWG 2016 Working Document no. 19.

Retzel A. 2016. Greenland Shrimp and Fish survey results for Atlantic cod in NAFO subareas 1A-1E (West Greenland) in 2015. ICES NWWG 2016 Working Document no. 20.

Fock H., Stransky C. and Bernreuther. 2016. Abundance for Sebastes norvegicus L., deep sea S. mentella and juvenile redfish (Sebastes spp.) off Greenland based on groundfish surveys 1985-2014. ICES NWWG 2016 Working Document no. 21.

Fock H. 2016. Update of Groundfish Survey Results for the Atlantic Cod Greenland offshore component After re-stratification of the survey 1982-2015. ICES NWWG 2016 Working Document no. 22.

Fock H. 2016. An age4plus index for the Atlantic Cod Greenland offshore component Based on recommendations from WKICE 2015 1989-2015. ICES NWWG 2016 Working Document no. 23.

Steingrund P. 2016. The biomass of Faroe Plateau cod 1860-1905 estimated by catch per unit effort and stock assessment data. ICES NWWG 2016 Working Document no. 24.

Steingrund P. 2016. The biomass of Faroe Plateau cod 1710-1859 estimated by qualitative cod years, dried cod export, and stock assessment data. ICES NWWG 2016 Working Document no. 25.

Vasilyev D. and Bobyrev A. 2016. Sebastes mentella fishery in the Irminger Sea: current state and perspectives. ICES NWWG 2016 Working Document no. 26.

Óskarsson G.J. and Guðmundsdóttir Á. 2016. Determination of F lim for Icelandic summerspawning herring (Her-Vasu). ICES NWWG 2016 Working Document no. 27.


[^0]:    * Asterisk indicates missing value(s).

[^1]:    Input units are thousands and kg - output in tonnes

[^2]:    ${ }^{1}$ Provisional data
    ${ }^{2}$ Based on estimates by observers onboard vessels

[^3]:    ${ }^{1)}$ Provisional

[^4]:    1) Provisional
