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# Report of the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK) 

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# International Council for the Exploration of the Sea Conseil International pour l'Exploration de la Mer 

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## Executive Summary

The ICES Working Group for the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK) met 26 April-5 May 2015 at Thuenen Institute in Hamburg, Germany. There were 24 participants (+ one by correspondence) from 9 countries. The main terms of reference for the Working Group were: to update, quality check and report relevant data for the working group, to update and audit the assessment and forecasts of the stocks, to produce a first draft of the advice on the fish stocks and to prepare planning for benchmarks next year. Ecosystem changes have been analytically considered in the assessments for cod, haddock and whiting in the form of varying natural mortalities estimated by the ICES Working Group on Multi Species Assessment Methods (WGSAM).

## Working procedures

WGNSSK has to deal with an increasing number of data limited stocks, formerly assessed by WGNEW. This increased substantially the workload. Therefore the working group met for 10 days instead of 8 days.

The progress initiated in late 2011 on the quality of the data collection process were pursued in 2016. Notably, a number of improvements were brought to the InterCatch data raising process in the last years, both through technical improvements of the InterCatch procedures and interface, and through the development of a number of exploratory scripts allowing rapid data screening and visualisation. Data were requested through a joint DCF-based data call for all assessment working groups, and the deadline for early data delivery were largely held.

The principle analytical models used for the stock assessments were SAM, XSA, TSA and the Aarts and Poos model (AAP), as well as SURBAR (for one data-limited stock).

WGNSSK works in close cooperation with WGMIXFISH and assessment and forecast results are directly used by WGMIXFISH to produce mixed fisheries advice. Similar links are established between WGNSSK and WGSAM to allow for an effective exchange of data and knowledge regarding multi species assessments.

## Benchmarks in 2016

Saithe in 4, 3.a and 6 as well as dab in 4 and 3.a has been benchmarked in 2016. Major changes were made to the saithe assessment (change from XSA to SAM, replacement of three national CPUE indices by a combined fishable biomass index, usage of ages 6 to 8 from the $3^{\text {rd }}$ quarter IBTS index). These changes led to a higher estimate of current SSB. Fishing mortality is still below the updated FMSY. For dab, an age based assessment with SURBAR could be provided for the first time and has been accepted as category 3 assessment.

An Interbenchmark was carried out for whiting in 4 and 7.d. Because of updated natural mortality estimates, the EU-Norway management strategy (fixed F without $\mathrm{B}_{\text {trigger }}$ and TAC constraints) used in previous years' advice is no longer considered precautionary. Following ACOM guidelines, an $\mathrm{F}=0.15$ when applied as part of the MSY approach (with $B_{\text {trigger }}$ ) leads to $<5 \%$ probability of falling below $\mathrm{Blim}_{\text {lim. }}$ In general, the stock dynamics of North Sea whiting are largely driven by recruitment and natural mortality and alternative management strategies should be evaluated for this stock.

WGNSSK rejected the update assessment for haddock in 4, 3.a and 6.a. An Interbenchmark will be carried out over the summer and advice is postponed to the autumn.

## State of the Stocks

The main impression in 2016 is that fishing mortality has been reduced significantly for many North Sea stocks of roundfish and flatfish compared to the beginning of the century. All fish stocks with agreed biomass reference points are above Blim, and only the SSB of cod in 4,3.a and 7.d and sole in 7.d is below MSY B trigger at the beginning of 2016. The assessment of haddock has been rejected by WGNSSK and stock status will be re-evaluated in autumn after an Interbenchmark. Several North Sea stocks are exploited around $\mathrm{F}_{\text {msy }}$ levels. Exceptions are cod in 4,3.a and 7.d and sole in 7.d. An important feature is that recruitment still remains poor compared to historic average levels for most gadoids. For whiting, the most recent recruitments are estimated to be higher but still do not reach historically observed high recruitment levels; however, the decreasing trend in SSB has stopped and the whiting stock has increased in recent years. The stock status of saithe has been revised during the 2016 benchmark process. According to the latest assessment, the saithe stock has fluctuated without trend above Btrigger since 1997. The southern North Sea and eastern channel areas are currently experiencing a steep increase of the plaice stock, with SSB values higher than ever observed in the assessment time series.

WGNSSK is also responsible for the assessment of several flatfish species that are mainly by catch in demersal fisheries. For all of these stocks, catch advice was provided in 2015 for the first time. The trends in abundance/biomass indices is decreasing for some of these stocks and increasing for others. In 2016 it was only necessary to determine whether the perception of the stock has changed compared to 2015; because perceptions have not changed compared to 2015, no reopening of the advice was needed.
The stock status of Nephrops stocks differs between functional units but globally a decrease compared to earlier years is observed for Nephrops in the North Sea.
The summary of stock status is as follows:

1) Nephrops: Abundance globally decreased compared to former years in the North Sea. The abundance of Nephrops in FU 6 and 7 has decreased further while the abundance in FU 8 and FU 9 has been stable in recent years. The abundance of Nephrops in 3.a is stable and at a high level. The 2015 harvest rates were in accordance with FMSY in FU 7, FU 9 and in 3.a. In contrast, harvest rates in FU 6 exceeded Fmsy in 2015. For FU 8 the harvest rate was very close to Fmsy. Because the TAC is set for the whole North Sea and not at a functional unit level, this contributes to $\mathrm{F}_{\text {MSY }}$ reference points being exceeded in some FUs. The FUs 5, 32, 33 and 34 are data limited. For these FUs, densities have to be assumed, along with other variables (e.g. discard rates, mean weights) and/or TV-survey estimates are uncertain. Given these assumptions, current harvest rates seem not to be problematic, apart from those for FU5.
2 ) Norway Pout: Will be assessed in autumn
3 ) Cod in area 4, 3.a. 20 and 7.d: Fishing mortality (F) has been declining since 2000 and is estimated to be above FMSY. Spawning-stock biomass (SSB) has increased from the historical low in 2006 and close to MSY Btrigger. Recruitment since 1998 remains poor.

4 ) Haddock in 3.a.20, 4 and 7.d: Assessment has been rejected. Will be re-evaluated in the autumn after an Interbenchmark

5 ) Whiting in 4 and 7.d: Spawning-stock biomass (SSB) has fluctuated around MSY $B_{\text {trigger. }}$ Fishing mortality ( F ) has been above Fmsy throughout the timeseries. Recruitment (R) has been low since 2003, with recruitment in 2014 and 2015 above previous years.
6 ) Saithe in 3.a, 4 and 6: Recruitment (R) has generally been below the longterm average since 2008. Fishing mortality (F) has been below Fmsy since 2013. Spawning-stock biomass (SSB) has fluctuated without trend, above MSY Btrigger since 1997.
7 ) Plaice in 4 and 3.a.20: The combined North Sea and Skagerrak stock is well above MSY Btrigger, increased in the past ten years, and has been at a record high for the last 5 years. Recruitment has been around the long-term average since the mid-90s. In recent years, fishing mortality ( F ) has been estimated around Fmsy.

8 ) Sole in 4: The spawning stock biomass (SSB) has increased since 2007 and has been estimated to be above MSY B trigger since 2012. Fishing mortality (F) has declined since 1997 and is estimated to be at FMSY in 2015. Recruitment (R) has fluctuated without trend since the early 1990s

9 ) Plaice in 7.d: Fishing mortality ( F ) has declined since the mid-1990s and has been below Fmsy since 2009. Spawning-stock biomass (SSB) has increased since 2008 and has been above MSY Btrigger since 2009. Recruitment (R) has been high since 2009.
10 ) Sole in 7.d: The spawning-stock biomass (SSB) has fluctuated without trend and is predicted to drop below MSY $\mathrm{B}_{\text {triger }}$ in 2016. Fishing mortality (F) has always been above Fmsy and increased over the years 2012-2015. Recruitment has been fluctuating without trend, and was among the lowest of the time series in 2012-2014, which has resulted in the decrease in recent SSB.
11 ) Category 3-6 stocks: In 2016 a new advice has been produced for Pollack in 4 and 3.a as well as grey gurnard in $4,7 . \mathrm{d}$ and 3.a. Pollack is a category 5 stock. However, information available suggest that the stock is at a low level compared to historic times. The time series of mature biomass index of grey gurnard from the IBTS-Q1 survey showed a strong increase from the beginning of $90^{\prime}$ 's, and has since fluctuated at a high level. Advice on other category 3 stocks will be updated next year. New survey information has not changed the perception of the stocks.

### 1.1 Terms of Reference

## Generic TORs

2015/2/ACOM05 The following ToRs apply to: AFWG, HAWG, NWWG, NIPAG, WGWIDE, WGBAST, WGBFAS, WGNSSK, WGCSE, WGDEEP, WGBIE, WGEEL, WGEF, WGHANSA and WGNAS.

## The working group should focus on:

a ) Consider and comment on ecosystem overviews where available;
b) For the fisheries relevant to the working group consider and comment on:
i) descriptions of ecosystem impacts of fisheries where available
ii ) descriptions of developments and recent changes to the fisheries
iii ) mixed fisheries overview, and
iv ) emerging issues of relevance for the management of the fisheries;
c ) Conduct an assessment to update advice on the stock(s) using the method (analytical, forecast or trends indicators) as described in the stock annex and produce a brief report of the work carried out regarding the stock, summarising where the item is relevant:
i) Input data (including information from the fishing industry and NGO that is pertinent to the assessments and projections);
ii ) Where misreporting of catches is significant, provide qualitative and where possible quantitative information and describe the methods used to obtain the information;
iii ) For relevant stocks estimate the percentage of the total catch that has been taken in the NEAFC Regulatory Area by year in the recent three years.
iv ) The developments in spawning stock biomass, total stock biomass, fishing mortality, catches (wanted and unwanted landings and discards) using the method described in the stock annex;
v ) The state of the stocks against relevant reference points;
vi ) Catch options for next year;
vii )Historical performance of the assessment and catch options and brief description of quality issues with these;
d ) Produce a first draft of the advice on the fish stocks and fisheries under considerations according to ACOM guidelines.

The working group is furthermore requested to:
e ) Consider and propose stocks to be benchmarked;
f) Review progress on benchmark processes of relevance to the expert group;
g ) Propose specific actions to be taken to improve the quality and transmission of the data (including improvements in data collection);
h) Prepare the data calls for the next year update assessment and for the planned data evaluation workshops;
i) Update, quality check and report relevant data for the stock:
i) Load fisheries data on effort and catches into the INTERCATCH database by fisheries/fleets;
ii ) Abundance survey results;
iii ) Environmental drivers.
j) Produce an overview of the sampling activities on a national basis based on the INTERCATCH database or, where relevant, the regional database.
k ) Identify research needs of relevance for the expert group.

## Specific TORs

2015/2/ACOM14 The Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), chaired by Alexander Kempf, Germany, and Jose De Oliveira*, UK, in meet in Hamburg, 26 April-5 May 2016 and by correspondence in September 2016 to:
a ) Address generic ToRs for Regional and Species Working Groups. The Norway pout assessments shall be developed by correspondence;
b) Check the relevance of the reopening procedure and report on reopened advice if appropriate.
Material and data relevant for the meeting must be available to the group no later than 12 April 2016 according to the Data Call 2016.

WGNSSK will report by 12 May 2016, and by 28 September 2016 (Norway pout) for the attention of ACOM. Concerning ToR b) the group will report on the ACOM guidelines on reopening procedure of the advice before 12 October and will report on reopened advice before 28 October.

### 1.2 InterCatch

### 1.2.1 Metier-based data call for WGNSSK (and other working groups)

The year 2012 represented a major change in the process of data collection for WGNSSK. Following an initiative launched by ICES WGMIXFISH in August 2011, it had been decided to merge the data calls and data collection of both groups WGNSSK and WGMIXFISH, on the basis of

1 ) improving the availability of metier-based data and their consistency with the stock-based data used for single-stock assessment.
2 ) allowing WGMIXFISH to meet earlier and as such integrate the mixed-fisheries advice within the single-stocks advice sheets.

In 2014 data limited stocks were included in the data call for the first time to improve the knowledge base for these stocks. Under the landing obligation these stocks become more important and discard information is a prerequisite to give catch advice and to carry out mixed fisheries scenarios under the landing obligation. In 2015, for the first time a joint data call for all relevant assessment working groups was launched.

The principle of the data call is to define the aggregation (metier) level for the data individual countries should deliver following the requirements of the EU Data Collection Framework (DCF), and to use these as the basis for providing and subsequently raising data for all North Sea demersal stocks. ICES InterCatch database was chosen as
the most appropriate tool to use until the planned Regional Data Bases are fully established and operational. Basic strata for the submission of catch and effort data were by country, quarter, area, metier and catch category.
In 2016, despite a new format of the data call the procedure for data submission was practically the same as in 2015. An official data call was issued by ICES, with a deadline for data delivery by 12th April 2016. No major issues occurred at that stage, and only few data were delayed compared to this date and some errors needed to be corrected before the working group without having a major impact on the work.

### 1.2.2 Data raising and allocation to unsampled strata

Major changes occured in recent years with the raising of data within InterCatch. Different initiatives can be mentioned here.

## 3 ) Technical improvements in the InterCatch interface:

- Allocation Group Setup: Define a group of unsampled catch/strata for which each distribution will be calculated according to the (for the group) allocated sampled catches/strata
- Automatic allocation 'same' strata: Automatically find and allocate identically sampled strata from other countries to unsampled catches/strata (with the identical stratum)
- Discard Group setup: Define a group of raised discards for which each discard weight will be calculated according to the (for the group) selected land-ing-discard ratios
- CATON and age/length data overviews: it is possible to examin all imported data in detail
- Allocation overview for pivot table/matrix: All unsampled strata are shown in the first column and all sampled strata are shown as the first row, then all the selected combinations are shown in the matrix.
- Possibility to save allocation schemes

4 ) Summary outputs and inspection of data before raising: the new features included in InterCatch allowed improved inspection and visualization of the data submitted by national data providers and a comparison with data from previous years. A generic R script has been developed in 2016 by Y. Vermard (IFREMER) mapping out the raw data, through e.g. quantification of the proportion of catches covered by sampling, identification of major gaps and outliers, plot of the age distribution and discards ratio of the various strata etc.
5 ) Raising procedures. Based on statistical principles discussed within WKPICS, RCMs, PGCCDBS and DC-MAP etc, the suggestions for the basis on which to proceed regarding raising of age distributions and discards ratio have been revisited. In 2012, the raising and allocating was based on finding similar strata from other countries, but this was judged not fully defendable in terms of statistical integrity. In 2016 the underlying principles applied were thus:
a) main strata are supposed to be sampled. In essence one should expect that the largest share of catches should have age-based and discards information in InterCatch. Which means that there is indeed a great number of unsampled strata (as we saw this last year), but in reality they
should represent only a minor part of the catches. Large strata without sampling information would need to be investigated further.
b ) Therefore, the suggestion was that by default, unsampled strata should be raised by all sampled strata, unless there is a good and informed reason for chosing differently after the data inspection process. Each stock coordinator developed general principles for the allocation scheme. The main principles are mentioned in the the respective report sections.

Ultimately, all these changes have triggered in-depth investigation and understanding of the data submitted, and are hopefully contributing to improved consistency and transparency in the assessment data. However, if more than one year needs to be raised the Intercatch procedure is still very time consuming. The saving of allocations schemes does not always function, especially when the metiers differ between years, and currently, only the age allocation scheme can be copied (not the discard ratio allocation scheme). It would be beneficial to allow for more flexible automatic matching based on e.g., gear type or area only. Also the possibility of entering allocation schemes via scripts (instead of the need to click through the options and metiers) would allow for fast sensitivity checks and would make Intercatch much more user-friendly.
Because of the landing obligation new catch categories need to be reported from 2016 onwards. There is currently little guidance about how the different catch categories should be reported, and how they should be used in raising discards where discard information is not provided. This is an issue that affects all Expert Groups that have to provide catch advice, so a common approach is needed. BMS landings, observer discards and log-book recorded discards should sum up to discard data provided so far (i.e. double-counting should be avoided), and when performing raising procedures, these three categories should be combined to provide raising factors, and the raising procedure in Intercatch should be adapted as necessary. This provides a robust approach, independent of how countries categorize catches when providing catch data. WGNSSK recommends that ICES provides a harmonized approach across all Expert Groups.

InterCatch summary data have been made available on the SharePoint, and will be investigated further during ICES WGMIXFISH.

By the end of the WG, the status of InterCatch use was as follows :

| Stocks for which data are imported | - Datar - Extracted (work started in InterCatc - | Exported (work finished in InterCatch - | Status of Data filled in |
| :---: | :---: | :---: | :---: |
| bll-nsea | 2015 Extracted | Exported | DataNOTusedForAssessment |
| cod-347d | 2015 Extracted | Exported | DataUsedForAssessment |
| dab-nsea | 2015 Extracted | Exported | Notfilled |
| fle-nsea | 2015 Extracted | No | Notfilled |
| gug-347d | 2015 Extracted | No | DataNOTusedForAssessment |
| had-346a | 2015 Extracted | Exported | DataUsedForAssessment |
| lem-nsea | 2015 Extracted | Exported | DataUsedForAssessment |
| mur-347d | 2015 Extracted | Exported | DataUsedForAssessment |
| nep-10 | 2015 Extracted | No | DataUsedForAssessment |
| nep-32 | 2015 Extracted | No | DataNOTusedForAssessment |
| nep-33 | 2015 Extracted | No | Notfilled |
| nep-34 | 2015 Extracted | Exported | DataUsedForAssessment |
| nep-3-4 | 2015 Extracted | No | DataNOTusedForAssessment |
| nep-5 | 2015 Extracted | Exported | DataUsedForAssessment |
| nep-6 | 2015 Extracted | Exported | DataUsedForAssessment |
| nep-7 | 2015 Extracted | Exported | DataUsedForAssessment |
| nep-8 | 2015 Extracted | Exported | DataUsedForAssessment |
| nep-9 | 2015 Extracted | Exported | DataUsedForAssessment |
| nep-IVnotFU | 2015 Extracted | No | Notfilled |
| nop-34 | 2015 Extracted | Exported | Notfilled |
| ple-eche | 2015 Extracted | Exported | DataNOTusedForAssessment |
| ple-nsea | 2015 Extracted | Exported | Notfilled |
| ple-skag | 2015 Extracted | Exported | DataUsedForAssessment |
| pol-nsea | 2015 Extracted | No | Notfilled |
| sai-3a46 | 2015 Extracted | Exported | DataUsedForAssessment |
| sol-eche | 2015 Extracted | Exported | DataUsedForAssessment |
| sol-nsea | 2015 Extracted | Exported | DataUsedForAssessment |
| tur-kask | 2015 Extracted | No | DataUsedForAssessment |
| tur-nsea | 2015 Extracted | Exported | DataUsedForAssessment |
| whg-47d | 2015 Extracted | Exported | DataUsedForAssessment |
| whg-kask | 2015 Extracted | Exported | DataUsedForAssessment |
| wit-nsea | 2015 Extracted | Exported | DataUsedForAssessment |

### 1.3 General uncertainty considerations

Data or inputs used in this report are based on sampling or on census. Typical census data are landings data from saleslips representing total landing, while sampled data are random samples (design based) used to produce estimates of total, relative indices or to characterize composition (like catch at age). All sources of input may introduce error in estimates/calculations and is a limiting factor in the amount of information and/or interpretation of model results. The scientist at this working group are only responsible for a modest fraction of the input data used and are relying heavily on assumptions regarding their validity and quality. The information based on sampling will contain sampling errors (random errors due to the stochastic nature of such sampling) and estimates of sampling error are generally not used by this working group. Such errors will show up in residuals (residual plots are an important diagnostic in the report), but other sources of error will also show up in the same residuals and are not easily separated from random errors. Non random errors are either bias or model errors. Systematic bias over time is a particular concern and an example of such can be underreporting of catches which will compromise the validity of the model results as basis for advice. Model errors may represent the use of the "wrong" equations to describe relations, but will in this report typically be linked to assumptions regarding natural mortality, the relationship between survey indices and stock size (catchability) and exploitation pattern. Some assumptions are needed since the Baranov and catch equations does not have unique solutions (too many parameters to estimate).
Assessment working groups are in many ways end users of data and it would be preferable to have such information presented as point estimates together with estimates of uncertainty or confidence bands and with a description of potential sources of bias and qualitative remarks related to specific observations. Intercatch is still not fully operational in this respect.

The working group appreciates the effort made by so many supporting hands involved in creating all information needed in fish stock assessment and is dependent on the
quality of information being upheld over time. An assessment working group is where information from the commercial fishery is handled together with fishery independent information to create estimates of stock status and the impact of fishing.
Demersal trawl surveys are the most used source of fishery independent information in this working group (WGNSSK). A demersal trawl survey uses a standardized procedure of trawling to create samples from a fish population. The "population" in statistical terms is the population of possible trawl stations with trawl station being the primary sampling unit. The estimates of uncertainty from a demersal trawl survey is very much dependent on the number of samples (trawl stations) and it seems that demersal trawl surveys on gadoid produces very similar estimates of uncertainty given the same number of trawl stations (ICES 1992) regardless of the size of the area. The relationship between sample size and precision can be illustrated using the following example: If a survey of 400 trawl stations produces an estimate (for a parameter of interest) with a corresponding relative standard error of 0.1 a reduction in survey effort to 100 trawl stations is likely to produce estimates with a relative standard error of 0.2 (divide the number of stations by 4 and the relative standard error is doubled). This is also likely to hold (at least as a rule of thumb) if one looks at results from a subarea of the original ( 400 station) area. When estimates of relative standard error approaches 0.3 , trends over time will be very difficult to detect, and with relative standard errors above 0.3 , the estimator can only be used to detect sudden events. WGNSSK recommends that along with survey index point estimates, DATRAS should also provide the uncertainty around these estimates as standard output.

### 1.4 Survey corrections during 2015 and 2016

During the last year no major resubmissions occurred for IBTS data and the indices produced by DATRAs as tuning indices only showed very minor changes compared to last year.
In contrast, the French and UK beam trawl surveys in area 7.d have been corrected and made more consistent between years. These corrections were related both to a modernization of calculation procedures (e.g. from Excel spreadsheets to R-scripts), and a more consistent use of data (e.g. use of a consistent set of prime stations, cleaning up an errors found, and harmonization of calculation methods with the way samples were collected, e.g. for age-length keys). Working documents are available under Annex 08 that describe in more detail the work done by national institutes. The final impact on the results of the plaice and sole in $7 . \mathrm{d}$ assessments is limited (see respective stock sections in this report).

### 1.5 Internal auditing and external reviews

ICES removed in general the external review process that had been in place for some years, and replaced it by an internal audit process within the Working Group itself. WGNSSK understands the motivations and reasoning behind this choice, and recognizes also that the process has certainly some merits and direct benefits such as increased participation and collaboration within the group across countries and stocks. However, WGNSSK wishes to underline that this audit is another heavy task pending on group members, and it was not possible to accomplish all audits during the meeting itself. WGNSSK operates with seldom more than one scientist per stock (sometimes even one scientist is responsible for two or more stocks), and there was simply not always enough time to have the reports finalized in time in order to carry out their
subsequent audit within the WG meeting. Audits had to be conducted by correspondence after the WG time, which is neither very efficient nor very motivating given the heavy workload most members usually operate with back in home institutes. WGNSSK recommends to explore alternative ways to strengthen the audit process (e.g., professional auditors as part of the ICES secretariat, external audit).

Finally, all WGNSSK stocks with an updated advice in 2016 could be covered by the internal audit (Table 1.5.1). The audit sheets can be found on the WGNSSK SharePoint.

Table 1.5.1. Fish stocks covered by the internal audit and external reviews:

| Fish Stock | Internal audit | External review |
| :---: | :---: | :---: |
| cod-347d | x |  |
| had-34 | Advice postponed | x |
| nep-5 | x |  |
| nep-6 | X |  |
| nep-7 | x |  |
| nep-8 | X |  |
| nep-9 | x |  |
| nep-10 | X |  |
| nep-32 | x |  |
| nep-33 | x |  |
| nep-34 | x |  |
| nep-iiia | x |  |
| nop-34 | no assessment in spring |  |
| ple-eche | x |  |
| ple-nsea | x |  |
| sai-3a46 | X | x |
| sol-eche | x |  |
| sol-nsea | X |  |
| whg-47d | x |  |
| whg-kask | no new advice in 2016 |  |
| bll-nsea | no new advice in 2016 |  |
| dab-nsea | no new advice in 2016 |  |
| fle-nsea | no new advice in 2016 |  |
| lem-nsea | no new advice in 2016 |  |
| wit-nsea | no new advice in 2016 |  |
| Tur-nsea | no new advice in 2016 |  |
| Tur-kask | no new advice in 2016 |  |
| Mur-347d | no new advice in 2016 |  |
| POL | X |  |
| GUR | x |  |

### 1.6 Mixed Fisheries

The mixed fisheries analyses have not been performed by WGNSSK over the last years. Instead, these are now being performed within the Working Group for Mixed Fisheries Advice for the North Sea (WGMIXFISH), which aims at evaluating the consistency of the ICES advice for the individual stocks in a mixed fisheries context, using the Fcube model (Ulrich et al., 2011).

The two groups have developed and issued a common data call since 2012, which greatly improved the quality and scheduling of data delivery. WGMIXFISH will meet in late May 2016 in order to integrate mixed-fisheries advice for the North Sea into single stock advice.

It is therefore referred to the ICES WGMIXFISH 2016 report for any further description of mixed-fisheries context.

However, the group discussed mixed fisheries issues under the landing obligation in the last years. There is a potential problem with choke species in the North Sea. Target species as well as bycatch species can become choke species for certain fleet segments. One way to deal with the situation could be to use the recently defined ranges for $\mathrm{F}_{\text {msy }}$ instead of point estimates (see ICES WKMSYREF III and IV, 2014-16). Ranges can introduce the flexibility needed to minimize the discrepancies in available quotas for species in a mixed fishery. However, further objectives are needed to determine which F inside the range should be applied. Ideas exist for an algorithm that minimises the discrepancy between TACs and ensures that a maximum of TACs can be fished out. It should be avoided that TACs are blindly set at the upper range for all stocks by managers. In the long-term there is no gain to fish stocks above $\mathrm{F}_{\text {msy }}$ as the yield becomes lower and the risk for the stocks increases. Selectivity in mixed fisheries should be improved instead to avoid choke effects.

The management of by-catch species (e.g., dab, lemon sole, turbot) by TAC further complicates the situation. If the TAC management for these species continues and Fmsy proxies will be implemented, these species can become serious choke species. The inter-institutional task force on multi annual plans between the European parliament, the council and the Commission write in their agreement (EU 8529/14): "With regard to by-catch species, the Co-legislators will have to determine, taking account of the available scientific advice, whether these are sufficiently covered through the management measures according to MSY for the key species". Policy has to define what sustainable exploitation means for bycatch species and it has to be evaluated by science whether MSY tagets for target stocks are enough to ensure a sustainable exploitation of bycatch species.

### 1.7 Multi species considerations

ICES gave advice on multi species considerations for the North Sea in 2013 for the first time to start a dialogue between ICES and its stakeholders on this topic. Simulations were carried out with the stochastic multi species model SMS to analyse $\mathrm{F}_{\text {msy }}$ in a multi species context. The multi species considerations can be found under: http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2013/2013/multNS.pdf

WGNSSK supports this step. However, the group also raised concerns about the data basis for the simulations (stomach data mainly from 1981 and 1991) and the high number of assumptions behind the model results.

Already in 2013 the group discussed the progress achieved under various initiatives such as ICES WGSAM $(2011,2012)$, ICES WKM-TRADE (2012) and the EU project MYFISH. The group noted that a multispecies benchmark as in the Baltic may be needed where the North Sea SMS model and keyrun settings are reviewed by external experts before a final multi species advice can be given.

There are many direct and indirect interactions between species, making it difficult to reach a single and robust best solution. Optimization scenarios carried out so far show that the result (target F) depends very much on the objectives (objective function) and SSB constraints used. The exact combination of species target F depends also on the weighting factors (e.g. price per kg when optimizing value) actually used for calculating these objectives. During a stakeholder workshop organized by ICES and MYFISH (ICES WKM-TRADE 2012) it has been agreed that when offering trade-offs, ICES can provide scenarios below $\mathrm{F}_{\text {MSY }}$ for the exploitation of some populations. This will allow a policy choice to be made within the limits defined and explained by ICES. FMSY rnages
(see also under mixed fisheries) could also help here to reach consenus based on a pretty good yield concept instead of trying to reach the absolute maximum for each stock, which is impossible given the biological interactions between predator and prey.

### 1.8 Frequency of assessments

ACOM provided for the first time criteria to test whether a category 1 stock could be a candidate for biennial assessments. The criteria are summarized in table 1.8.1.

Table 1.8.1 Criteria to be applied to identify candidate stocks for less frequent assessment.

| Stock Category | Criteria to be used to identify candidate stocks for less frequent |
| :---: | :---: |
| assessment. |  |

Cat. 1 and 2
Stocks are considered candidates for biennial assessment if:
The advice for the stock has been 0 -catch or equivalent for the latest three advice years.

Stocks are considered candidates for biennial assessment if the following criteria are fulfilled simultaneously:
Life span (i.e. maximum normal age) of the species is larger than 5 years
The stock status in relation to the reference points is according to the MSY criteria F(latest assessment year) <= $1.1 \times$ Fmsy OR if Fmsy range has been defined: F (latest assessment year) is <= Fupper (upper bound in F range) AND SSB(start of intermediate year) $>=$ MSY Btrigger

The average contribution to the catch in numbers of the recruiting year class in latest 5 years is less than $25 \%$ of the total catch in numbers. Should be calculated as the average over the latest five years of the catch in numbers of first age divided by the total catch in number by year.

The retrospective pattern, based on a seven years peel of Mohn's Rho index, shows that F is consistently underestimated by less than $20 \%$

The formula to be used in the calculations is:
$\rho=\frac{1}{7} \sum_{u=Y-7}^{Y-1}\left(1-\frac{F_{u, u}}{F_{u, Y}}\right)$. The result should be $<0.20$,
where $F_{u, u}$ is $F$ in year $u$ estimated from an assessment that ends in year $u$, and $F_{u, Y}$ is the F in year u estimated from the most recent assessment (which ends in year Y )

Cat. 3

Cat 4-5-6
By default all stocks in this category are considered candidates for biennial or triennial assessment.

By default all stocks in this category are considered candidates for triennial assessment.

Results of the criteria check for cateogry 1 stocks can be found in the respective stock sections. In conclusion, saithe, plaice in 4 and 7 .d as well as sole in 4 are currently candidates for biennial assessments based on the ACOM criteria alone (table 1.8.2). In general, only if the criteria based on the status of the stock are met, the other criteria need to be tested.

Table 1.8.2 Information on whether criteria for a biennial assessment are met.

| Fish Stock | Live span >= <br> 5 years | Stock status: Fcur <=1.1x Fmsy or inside F MSY ranges | Stock status: SSB (start of the intermediate year) >= MSY Btrigger | Contribution of first age to total catches in the last 5 years $\mathbf{<} \mathbf{2 5 \%}$ | Mohn's Rho index shows that $F$ is not consistently underestimated by less than $\mathbf{2 0 \%}$ | Potential <br> candidate <br> for less <br> frequent <br> assessments ? |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| cod-347d | Yes | Yes | No | No | Yes | No |
| had-346a* | Yes | Yes | No | Yes | No | No |
| ple-eche | Yes | Yes | Yes | Yes | Yes | Yes |
| ple-nsea | Yes | Yes | Yes | Yes | Yes | Yes |
| sai-3a46 | Yes | Yes | Yes | Yes | Yes | Yes |
| sol-eche | Yes | No | Yes | ? | ? | No |
| sol-nsea | Yes | Yes | Yes | Yes | Yes | Yes |
| whg-47d | Yes | No | Yes | No | Yes | No |

* Based on the rejected update assessment. Needs to be re-evaluated in autumn after the Interbenchmark!

Although saithe passes all criteria in table 1.8.2, WGNSSK argues that it still may not be a good candidate for less frequent assessments. The reason is that the stock assessment suffers from structural uncertainties. The input data for the assessment is uncertain (i.e. year effects in the IBTS q3 survey), as shown during the benchmark (WKNSEA 2016) and work following the benchmark. Minor modifications to the weighting of tuning series can greatly change perception of the stock. This, coupled with the uncertainties in the forecast (a substatial part of the adviced catch comes from year classes where year class strength needs to be assumed) questions a reduction in the frequency of assessments.

In general, WGNSSK members expressed during their discussion in 2015 that relatively simple tests would generally be insufficient to determine the risk of unwanted outcomes, should the frequency of assessments for a particular stock be reduced. Such an exercise requires a simulation analysis of the type used to evaluate management plans and strategies. An approach of this kind takes considerable time that is not available during the WG meeting.

The working group would also like to point out as in previous years that time spent at the yearly assessment meetings represent only a modest fraction of the work related to collect and prepare the information needed in these meetings. WGNSSK is of the opinion that "new information" is best assessed by using the information directly in an assessment model (update run), which is usually quick to produce once the necessary data have been put together.

As an important point, WGNSSK underlines that stock assessment and corresponding advice is an integrated part of long term management strategies and the management system in general. Any change in the frequency of assessments represents a change to the management strategy/system itself (or at least the implementation of it) and needs to be evaluated in the same way as management strategies in general (e.g., Management Strategy Evaluations, MSE).

### 1.9 Special requests

No larger special requests have been received during 2015 and beginning of 2016 for North Sea stocks.

### 1.10 References

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### 2.1 Introduction

The demersal fisheries in the North Sea can be categorised as a) human consumption fisheries, and $b$ ) industrial fisheries which land the majority of their catch for reduction purposes. Demersal human consumption fisheries usually either target a mixture of roundfish species (cod, haddock, whiting), a mixture of flatfish species (plaice and sole) with a by-catch of roundfish and other flatfish (e.g., turbot, brill dab), or Nephrops with a bycatch of roundfish and flatfish. A fishery directed at saithe with some bycatch of hake and other roundfish exists along the shelf edge.
The industrial fisheries which used to dominate the North Sea catch in weight have become much less prominent. Human consumption landings have steadily declined over the last 30 years, with an intermediate high in the early 80 's. The landings of the industrial fisheries show the largest annual variations, resulting from variable recruitment and the short life span of the main target species. The total demersal landings from the North Sea reached over 2 million $t$ in 1974, were around 1.5 million $t$ in the 1990s and are currently around 600000 t , of which over half is industrial fisheries.
For some stocks, the North Sea assessment area may also cover other regions adjacent to ICES Subarea 4. Thus, combined assessments are made for cod including 3.aN (Skagerrak) and 7d, for haddock including VIa and 3.a, Norway pout including 3.a, for whiting including 7d, and for saithe including 3.a and 6.a. Since the benchmark in 2015 also plaice in 3.a is now assessed together with plaice in area 4 . The state of Nephrops stocks are evaluated on the basis of discrete Functional Units (FU) on which estimates of appropriate removals are founded. Quota management for Nephrops is still carried out at the Subarea and Division level, however.

The sandeel assessment has been moved to ICES HAWG and is therefore not presented anymore in WGNSSK report.

The analysis of biological interactions (predator-prey relationships) among species has been a central theme in ICES over the last 30 years, primary for the Baltic Sea and the North Sea. The2011 and 2014 North Sea key run performed by the multispecies group WGSAM represents the ultimate state of the art in terms of multispecies assessment, with the dynamic estimation of predation mortality. This has led to the publication of the first multispecies advice by ICES in 2013 (http://www.ices.dk/sites/pub/Publica-tion\ Reports/Advice/2013/2013/mult-NS.pdf) . The single-stock assessments and advice presented in this report are not produced by the multispecies assessment model, but time variant values of natural mortalities estimated by multispecies assessments for cod, haddock, and whiting are incorporated in the assessments of these species. Flatfish are not part of the current multi species assessment and more work is needed to incorporate also information on flatfish in the multi species advice.
Gear types vary between fisheries. Human consumption fisheries use otter trawls, pair trawls, Nephrops trawls, seines, gill nets, or beam trawls, while industrial fisheries use small meshed otter trawls. Trends in reported effort in the major fleets fishing in the North Sea are described annually by the ICES WG on Mixed Fisheries Advice for the North Sea (ICES WGMIXFISH 2016), which meets straight after the WGNSSK. Both WG share a joint data call issued by ICES for fulfilling the data needs of both groups.
The data distinguish between two basic concepts, the Fleet (or fleet segment), and the Métier. Their definition has evolved with time, but the most recent official definitions
are those from the EC's Data Collection Framework (DCF, Reg. (EC) No 949/2008), which we adopt here:

- A Fleet segment is a group of vessels with the same length class and predominant fishing gear during the year. Vessels may have different fishing activities during the reference period, but might be classified in only one fleet segment.
- A Métier is a group of fishing operations targeting a similar (assemblage of) species, using similar gear, during the same period of the year and/or within the same area and which are characterized by a similar exploitation pattern.

Fleets and métiers were defined to match with the available economic data and the cod long term management plan. In 2013 and 2014 WGMIXFISH included new stocks in its analyses (plaice and sole in the Eastern Channel as full analytical stocks; hake in the North Sea and plaice in Skagerrak as additional "lpue" stocks as well as turbot, see WGMIXFISH 2013 and 2014 report). Plaice in the Skagerrak has been merged with plaice in 4 in 2015. In 2015 and 2016 WGMIXFISH focused again on the full analytical assessed stocks as well as Nephrops only but no longer on turbot and hake. This will most likely change in the near future again as a lot of the stocks without full analytical assessment (e.g., turbot, hake, witch, lemon sole) are important by-catch and can become potential "choke species" once under the landing obligation.
Ultimately, WGMIXFISH has identified 43 national demersal fleets (combination of country, main gear and vessel size category, plus an "others" OTH fleet) from nine countries. These fleets engage in one to four different métiers each, resulting in 118 demersal fisheries (country*fleet* métier*area) in the North Sea, Skagerrak and Eastern Channel.

ICES WGMIXFISH produces a number of synthetic Figures describing main trends, of effort and catches and landings by fleet and stock. The effort time series is not $100 \%$ consistent. It is planned to issue a data call and ask countries to resubmit a full time series of effort data. The following paragraph is based on data from WGMIXFISH 2015 and will be updated once a more consistent time series becomes available:

Overall nominal effort (kW-days) by EU demersal trawls regulated in the cod management (TR1, TR2, TR3, GN1, GT1, LL1,BT1,BT2) in the North Sea, Skagerrak, and Eastern Channel had been substantially reduced since the implementation of the two successive effort management plans in 2004 and 2008 ( $-30 \%$ between 2004 and 2014, $-12 \%$ between 2008 and 2014). Following the introduction of days-at-sea regulations in 2003, there was a substantial switch from the larger mesh ( $>100 \mathrm{~mm}$, TR1) gear to the smaller mesh (70-99 mm, TR2) gear. Subsequently, effort by TR1 has been relatively stable, whereas effort in TR2 and in small-mesh beam trawl ( $80-120 \mathrm{~mm}, \mathrm{BT} 2$ ) has shown a pronounced decline $(+2 \%,-43 \%$, and $-49 \%$, respectively, between 2004 and 2014). Gill- and trammelnet fisheries have increased ( $+20 \%,+13 \%$ ). Effort in largemeshed beam trawl ( $\geq 120 \mathrm{~mm}$, BT1) has increased significantly in 2012 and 2013 after a decade of continuous decline. Nominal effort reported by Norway has increased since 2011 due to the generalization of electronic logbooks. For 2014 the same effort had to be assumed as observed in 2013 because of data quality issues.

Technical interactions appear between stocks when they are caught by the same gear during a fishing operation (mixed fisheries). Ideally the technical interaction should then be studied at the scale of the fishing operation to prevent artificially creating technical interaction between stocks that might only be caught at day/night or in different areas/timing of the year. However, the finest available information is per
stock/gear/area/season from the WGMIXFISH data base, aggregating the catches at an already quite high level.
Knowing these limitations, this database can still be used to provide information on technical interactions. The methodology used here consists in computing the sum of catches per strata of one species given that a second species is also present in the total catches of this stratum. This value is then divided by the total catches of the first species:


Where $i$ and $j$ are the two species for which the technical interaction is assessed.
PresInLandings $s_{\text {strata, } j}$ is a Boolean and equals 0 if the total landings of species $j$ for a given strata is less than $5 \%$ the total landings for that strata and 1 otherwise to remove the by catch effect. Strata corresponds to the provided disaggregation of the landings in WGMIXFISH. The minimum level being Season=Quarter, Metier=Metier level 6, Area=ICES division.

The following table shows the technical interaction between stocks. Red cells indicate that the species are caught together to a large extent. Orange cells indicate less strong interactions while yellow cells indicate a weak interaction.


### 2.2 Main management regulations

The near-collapse of the North Sea cod stock in the beginning of the 2000s led to the introduction of effort restrictions alongside TACs as a management measure within EU fisheries. There has also been an increasing use of single-species multi-annual management plans, partly in relation to cod recovery, but also more generally. With the implementation of the landing obligation in 2016 mixed fisheries multi species plans are under development but have not yet been agreed

The management frames can be summarised as such:

### 2.2.1 Landings obligation

Fisheries in Norwegian waters have been subject to a landing obligation for cod and haddock from 1987 and for most species since 2009. A landing obligation for EU fisheries on demersal species in the North Sea is implemented from 2016 in a phased approach with all quota stocks subject to the landings obligation from 2019 onwards. Detailed definitions of the landing obligation can be found in Article 15 of regulation 1380/2013. For 2016 discard plans have been agreed defining for which species, gear and mesh size combinations the landing obligation applies in 2016. Table 2.2.1, 2.2.2 and 2.2.3 show these combinations relevant for WGNSSK stocks in areas 4, 3.a, 6a and 7 d . The discard plans will be amended to define which additional species and gear combinations will fall under the landing obligation in 2017 and 2018. Until 2019 it is expected that the landing obligation is fully implemented.

Table 2.2.1. Fisheries under the landing obligation in area 4 and 3.a (from Comission delegated regulation (EU) 2015/2440).

| Fishing gear ${ }^{19} 9$ | Mesh rize | Species concerned |
| :---: | :---: | :---: |
| Trawls: <br> OTB, OTT, OT, PTB, PT, TBN, TBS, OTM, PTM, TMS, TM, TX, SDN, SSC, SPR, TD, SX, SV | $>100 \mathrm{~mm}$ | All catches of saithe (if caught by a saithe targeting vessel (3), plaice and haddock. <br> All by-catches of Northern prawn. |
| Trawls: <br> OTB, OTT, OT, PTB, PT, TBN, TBS, OTM, ITM, TMS, TM, TX, SDN, SSC, SPR, TB, 5X, SV | In ICES Subarea IV and in Union waters of ICES Division Ha: $80-99 \mathrm{~mm}$ <br> In ICES Division Illa: 70-99 mm | In all areas, all catches of Norway lobster and common sole (). <br> All by-catches of Northern prawn. <br> In ICES Division Illas: all catches of haddock. |
| Traws: <br> OTB, OTT, OT, ITB, FT, TBN, TBS, OTM, PTM, TMS, TM, TX, SDN, SSC, SPR, TB, SX, SV | 32.69 mm | All catches of Northern prawn. |
| Beam trawls: <br> TBB | $>120 \mathrm{~mm}$ | All carches of plaice. <br> All by-catches of Northern prawn. |
| Beam trawls: TBB | $80-119 \mathrm{~mm}$ | All catches of common sole. <br> Any by-catcbes of Northern prawn. |
| Gillnets, trammel nets and entangling nets: <br> GN, GNS, GND, GNC, GTN, GTR, GEN, GNF |  | All catches of common sole. <br> All by-catches of Northern prawn. |
| Hooks and lines: <br> LLS, LLD, LL, LTL, LX, LHP, LHM |  | All catches of hake. <br> All by-catches of Northern prawn. |
| Traps: <br> FFO, HIX, FYK, IPN |  | All catches of Norway lobster. <br> All by-catches of Northern prawn. |

(1) Gear codes used in this Table refer to those codes in Amex XI to Commission implementing Regulation (FU) No 404/2011 haying down detailed rules for the implementation of Council Regulation ( EC ) No 1224/2009 establizhing a Cotumunity control system for ensuring compliance with the rules of the Common fisheries policy.
(7) For the vessels whose LOA is less than 10 metres gear codes used in this table refer to the codes from the FAO gear classification
(7) Vessels are considered as saithe targering if, when using trawls with mesh size $\geq 100 \mathrm{~mm}$, they have had annual average landings of saithe of $=505$ of all landings by the vessel taken in both EU and thind country zone of the North Sea over the period of x-4 to $x-2$ where $x$ is the year of application: i.e. 2012-2014 for 2016 and 2013 -2015 for 2017
(7) Except in ICFS division Illa when fishing with trawls with a mesh siat of at least 90 mm equipped with a sop panel of at least 270 mm mesh size (diamond mesh) or at leagt 140 mm mech zize (zquare mesh) or 120 mus squre meah panel placed 6 to 9 meters from the cod-end.

Table 2.2.2. Fisheries under the landing obligation in area 6a (from Comission delegated regulation (EU) 2015/2438).
(a) Fisheries in ICES division Vla and Union waters of ICES division Vb

| Fishery | Gear Code | Fuling gear description | Mesh Size | Landing Obligation |
| :---: | :---: | :---: | :---: | :---: |
| Cod (Gadus morlua), Haddock (Mdlanogramunus arglifinus), Whiting (Merlangius merlangus) and Saithe (Pollachius virens) | OTB, SSC, OTT, PTB, SDN, SPR, TBN, TBS, TB, SX, SV, OT, PT, TX | Trawls \& Seines | All | Where total landings per vessel of all species in 2013 and 2014 consist of more than $10 \%$ of the following gadoids: cod, haddock, whiting and saithe combined, the landing obligation shall apply to baddock. |
| Norway lobster (Neplinops norvegious) | OTB, SSC, OTT, PTB, SDN, SPR. FPO, TBN, TB, TBS, SX, SV, FLX, OT. PT. TX | Trawls, Seines, Pots, Traps \& Creels | All | Where the total landings per vessel of all species in 2013 and 2014 consist of more than $30 \%$ of Norway lobster, the landing obligation shall apply to Norway lobster. |

Table 2.2.3. Fisheries under the landing obligation in area 7 d (from Comission delegated regulation (EU) 2015/2438).

| Fishery | Gear Code | Fishing gear | Mesh Size | Landing Obligation |
| :---: | :---: | :---: | :---: | :---: |
| Common Sole (Solra soleat) | TBB | All Beam trawls | All | All catches of common sole are subject to the landing obligation. |
| Common Sole (Solea soled) | OTT, OTB, TBS, TBN, TB, PTB, OT. PT. TX | Trawls | $<100 \mathrm{~mm}$ | Where the total landings per vessel of all species in 2013 and 2014 consist of more than $5 \%$ of common sole, the landing obligation shall apply to common sole. |
| Common Sole (Solea solia) | GNS, GN, <br> GND, GNC, <br> GTN, GTR, <br> GEN | All Trammel nets \& Gill nets | All | All catches of common sole shall be subject to the landing obligation. |
| Cod (Gadus mornua), Haddock (Melanogrammus neglefinus). Whiting (Merlangius merlangus) and Saithe (Pollachius virens) | OTB, SSC, OTI, <br> ITB, SDN, SPR <br> TBN, TBS, TB, <br> SX. SV, OT, PT. <br> TX | Trawls and Scines | All | Where total landings per vessel of all species in 2013 and 2014 consist of more than $25 \%$ of the following gadoids: cod, haddock, whiting and saithe combined, the landing obligation shall apply to whiting. |

There is a high probability that the implementation of the EU landing obligation with its complex definitions, exemptions and rules (e.g. de-minimis, high survival, $9 \%$ interspecies flexibility) has implications for the quality of monitoring of the catches and the quality of assessments of the stock status and exploitation rate. De-minimis exemptions and the $9 \%$ inter-species flexibility rule can lead to serious implications for stocks dependent on the interpretation of the respective paragraphs in the regulation (STECF 2014a, b). The possibility to use up to $9 \%$ of the quota of a target species for bycatch of any other species constitutes a major factor for uncertainty in future management as it is not possible to predict what will happen at least in the first years.

In 2016 a high survival exemption has been granted for the main metiers catching Nephrops in 3.a and discarding of Nephrops below the minimum conservation reference size (MCRS) up to a deminins of $6 \%$ is still allowed in area 4 . Also the MCRS has been
reduced substantially in 3.a. WGNSSK tries to take this into account in the forecasts for Nephrops by assuming the 2015 selection pattern in the respective fisheries but that only discards below the agreed MCRS continue under the landing obligation.

Also for sole and haddock several de-minims exemptions have been agreed. The default ICES assumption is that the same exploitation patterns as observed in recent years will continue and former discards are now called unwanted catch. How much of this unwanted catch will be landed in the future (catch cateogry BMS) and how much will still be discarded is pure speculation. Given that stocks are impacted by the total F independent of how the total catch is split up (at least under the assumption of no survival of discards), the results of forecasts are robust to assumptions regarding which fraction of the total catch will be landed. In contrast, the landing obligation will mean a serious change and therefore exploitation patterns of fleets will most likely change in the future. Predicting these changes is impossible at the current stage what leads to an increased uncertainty in short term forecasts until more information becomes available (from 2017 onwards).

### 2.2.2 Effort limitations

For vessels registered in EU member states, effort restrictions in terms of days at sea were introduced in 2003 and subsequently revised annually. Initially days at sea allowances were defined by calendar month. From 2006 the limit was defined on an annual basis. The maximum number of days a fishing vessel could be absent from port varied according to gear type, mesh size (where applicable) and region. A complex system of 'special conditions' (SPECONs) developed upon request from the Member States, whereby vessels could qualify for extra days at sea if special conditions (specified in the Annexes) were met. Increasingly detailed micromanagement took place until 2008 (Ulrich et al., 2012). A detailed description of these categories as well as the corresponding days at sea can be found in STECF (2008).
In 2008 the system was radically redesigned. From 2009, a total effort limit (measured in kW days) is set and divided up between the various nation's fleet effort categories. The baselines assigned in 2009 were based on track record per fleet effort category averaged over 2004-2006 or 2005-2007 depending on national preference, and the effort ceilings were updated in 2010. After some reductions based on the cod management plan to support the recovery of the cod stock, an effort roll over was decided for 2013, 2014 and 2015.

The areas are Kattegat, the part of 3.a not covered by Skagerrak and Kattegat, ICES zone 4, EC waters of ICES zone 2a, ICES zone 7d, ICES zone 7.a, ICES zone 6a and EC waters of ICES zone 5 b . The grouping of fishing gear concerned are: Bottom trawls, Danish seines and similar gear, excluding beam trawls of mesh size: TR1 ( $\leq 100 \mathrm{~mm}$ ) TR2 ( $\leq 70$ and $<100 \mathrm{~mm}$ ) - TR3 ( $\leq 16$ and $<32 \mathrm{~mm}$ ); Beam trawl of mesh size: BT1 ( $\leq$ $120 \mathrm{~mm})-$ BT2 ( $\leq 80$ and $<120 \mathrm{~mm}$ ); Gill nets excluding trammel nets: GN1; Trammel nets: GT1 and Longlines: LL1. The respective effort limitations per area per gear can be found in annex IIa and Appendix 1 to Annex IIa in the annual TAC and quota regulations

Table 2.1.1 Maximum allowable fishing effort in kilo watt days in 2013, 2014, 2015 and 2016.

Geographical area: Skagerrak, that part of ICES division 3. a not covered by the Skagerrak and the Kattegat; ICES subarea 4 and EU waters of ICES division 2a; ICES division 7d

| Regulated |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TR1 | 895 | 3385928 | 954390 | 1409 | 533451 | 157 | 257266 | 172064 | 6185460 |
| TR2 | 193676 | 2841906 | 357193 | 0 | 6496811 | 10976 | 748027 | 604071 | 5127906 |
| TR3 | 0 | 2545009 | 257 | 0 | 101316 | 0 | 36617 | 1024 | 8482 |
| BT1 | 1427574 | 1157265 | 29271 | 0 | 0 | 0 | 999808 | 0 | 1739759 |
| BT2 | 5401395 | 79212 | 1375400 | 0 | 1202818 | 0 | 28307876 | 0 | 6116437 |
| GN | 163531 | 2307977 | 224484 | 0 | 342579 | 0 | 438664 | 74925 | 546303 |
| GT | 0 | 224124 | 467 | 0 | 4338315 | 0 | 0 | 48968 | 14004 |
| LL | 0 | 56312 | 0 | 245 | 125141 | 0 | 0 | 110468 | 134880 |

The STECF and ICES WGMIXFISH has performed annual monitoring of deployed effort trends since 2002. In addition, a more detailed overview and analyses of the various measures implemented in the frame of the cod recovery plan can be found in the 2011 joint STECF/ICES evaluation of this plan (ICES WKROUNDMP 2011, Kraak et al., 2013).

### 2.2.3 Stock-based management plans

Cod, saithe, whiting, plaice and sole are currently subject to multi-annual management strategies (the latter two, being EU strategies, not EU-Norway agreements). These plans all consist of harvest rules to derive annual TACs depending on the state of the stock relative to biomass reference points and target fishing mortality. The harvest rules also impose constraints on the annual percentage change in TAC. These plans have been discussed, evaluated and adopted on a stock-by-stock basis, involving different timing, procedures, stakeholders and scientists involved, disregarding mixedfisheries interactions (ICES WGMIXFISH 2012). The technical basis of the individual management plans is detailed in the relevant stock section. Most of these plans are no longer used as basis of advice and to set TACs due to benchmarks and the general change from individual target fishing mortalities to $\mathrm{F}_{\mathrm{msy}}$.

With the new CFP, the demand for mixed fisheries management plans covering all species caught in a fishery is increasing. However, so far no multi species (fishery based) management plans have been agreed for the greater North Sea. With the implementation of the landing obligation by 2016 for the North Sea demersal fisheries, problems caused by the management of mixed fisheries with single species plans will become more evident. In addition, benchmarks have caused major changes in the assessment and reference points in the last years.

### 2.2.4 Additional Technical measures

The national management measures with regard to the implementation of the available quota in the fisheries differ between species and countries. The industrial fisheries are subject to regulations for the by-catches of other species (e.g. herring, whiting, haddock, cod). Technical measures relevant to each stock are listed in each stock section. To these additional management measures belong e.g., real time closures or Fully Documented Fisheries (FDF).

### 2.2.4.1 Minimum landing size/ Minimum conservation reference size

"Undersized marine organisms must not be retained on board or be transhipped, landed, transported, stored, sold, displayed or offered for sale, but must be discarded immediately to the sea" (EC 850/98)). After the implementation of the landing obligation minimum landing sizes have been transformed into Minimum Conservation Reference Sizes (MCRS) that apply from 2016 onwards. The current MCRS can be found in table 2.2.4.1. Individuals below MCRS have to be landed now but are not allowed to be sold for human consumption.

Table 2.2.4.1. Current MCRS

|  | SPECIES | MCRS REGION 1-5 |
| :--- | :--- | :--- |

### 2.2.4.2 Minimum mesh size

Regulations on mesh sizes are more complex than those on landing sizes, as they differ depending on gears used, target species and fishing areas. Many other accompanying measures are implemented simultaneously with mesh sizes. They include regulations on gear dimensions (e.g. number of meshes on the circumference), square-meshed panels, and netting material. The most relevant mesh size regulations of EC No 2056/2001 are presented below.

## Towed nets excluding beam trawls

Since January 2002, the minimum mesh size for towed nets fishing for human consumption demersal species in the North Sea is 120 mm . There are however many derogations to this general rule, and the most important are given below:

- Nephrops fishing. It is possible to use a mesh size in range $70-99 \mathrm{~mm}$, provided catches retained on board consist of at least $30 \%$ of Nephrops. However, the net needs to be equipped with a 80 mm square-meshed panel if a mesh size of 70-99 mm is to be used in the North Sea and if a mesh size of 90 mm is to be used in the Skagerrak and Kattegat the codend has to be square meshed.
- Saithe fishing. It is possible to use a mesh size range of $110-119 \mathrm{~mm}$, provided catches consist of at least $70 \%$ of saithe and less than $3 \%$ of cod. This exception however does not apply to Norwegian waters, where the minimum mesh size for all human consumption fishing is 120 mm . Since January 2002 Norwegian trawlers (human consumption) have had a minimum mesh size of 120 mm in EU-waters. However, since August 2004 they have been allowed to use down to 110 mm mesh size in EU-waters (but minimum mesh size is still 120 mm in Norwegian waters).
- Fishing for other stocks. It is possible to use a mesh size range of 100-119 mm , provided the net is equipped with a square-meshed panel of at least 90 mm mesh size and the catch composition retained on board consists of no more than $3 \%$ of cod.
- 2002 exemption. In 2002 only, it was possible to use a mesh size range of $110-119 \mathrm{~mm}$, provided catches retained on board consist of at least $50 \%$ of a mixture of haddock, whiting, plaice sole, lemon sole, skates and anglerfish, and no more than $25 \%$ of cod.


## Beam trawls

- Northern North Sea. It is prohibited to use any beam trawl of mesh size range 32 to 119 mm in that part of ICES Subarea 4 to the north of $56^{\circ} 00^{\prime} \mathrm{N}$. However, it is permitted to use any beam trawl of mesh size range 100 to 119 mm within the area enclosed by the east coast of the United Kingdom between $55^{\circ} 00^{\prime} \mathrm{N}$ and $56^{\circ} 00^{\prime} \mathrm{N}$ and by straight lines sequentially joining the following geographical coordinates: a point on the east coast of the United Kingdom at $55^{\circ} 00^{\prime} \mathrm{N}, 55^{\circ} 00^{\prime} \mathrm{N} 05^{\circ} 00^{\prime} \mathrm{E}, 56^{\circ} 00^{\prime} \mathrm{N} 05^{\circ} 00^{\prime} \mathrm{E}$, a point on the east coast of the United Kingdom at $56^{\circ} 00^{\prime} \mathrm{N}$, provided that the catches taken within this area with such a fishing gear and retained on board consist of no more than $5 \%$ of cod.
- Southern North Sea. It is possible to fish for sole south of $56^{\circ} \mathrm{N}$ with $80-99$ mm meshes in the cod end, provided that at least $40 \%$ of the catch is sole, and no more than $5 \%$ of the catch is composed of cod, haddock and saithe.


## Combined nets

It is prohibited to simultaneously carry on board beam trawls of more than two of the mesh size ranges 32 to $99 \mathrm{~mm}, 100$ to 119 mm and equal to or greater than 120 mm .

## Fixed gears

The minimum mesh size of fixed gears is of 140 mm when targeting cod, which is when the proportion of cod catches retained exceeds $30 \%$ of total catches.

### 2.2.4.3 Closed areas

## Twelve mile zone

Beam trawling is not allowed in a 12 nm wide zone along the British coast, except for vessel having an engine power not exceeding 221 kW and an overall length of 24 m maximum. In the 12 mile zone extending from the French coast at $51^{\circ} \mathrm{N}$ to Hirtshals in Denmark trawling is not allowed to vessels over 8 m overall length. However, otter trawling is allowed to vessels of maximum 221 kW and 24 m overall length, provided that catches of plaice and sole do not exceed $5 \%$ of the total catch. Beam trawling is only allowed to vessels included in a list that has been drawn up for the purposes. The number of vessels on this list is bound to a maximum, but the vessels on it may be replaced by other ones, provided that their engine power does not exceed 221 kW and their overall length is 24 m maximum. Vessels on the list are allowed to fish within the twelve miles zone with beam trawls having an aggregate width of 9 m maximum. To this rule there is a further derogation for vessels having shrimping as their main occupation. Such vessels may be included in annually revised second list and are allowed to use beam trawls exceeding 9 m total width.

## Plaice box

To reduce the discarding of plaice in the nursery grounds along the continental coast of the North Sea, an area between $53^{\circ} \mathrm{N}$ and $57^{\circ} \mathrm{N}$ has been closed to fishing for trawlers with engine power of more than $221 \mathrm{kw}(300 \mathrm{hp})$ in the second and third quarter since 1989, and for the whole year since 1995. Beare et al. (2013) conducted a thorough analysis of the potential effect of the plaice box on the stock of plaice, and concluded that no significant effect, neither positive nor negative, could be related to the implementation of the plaice box.

## Sandeel box

In the light of studies linking low sandeel availability to poor breeding success of kittiwake, ICES advised in 2000 for a closure of the sandeel fisheries in the Firth of Forth area east of Scotland. All commercial fishing was excluded, except for a maximum of 10 boat days in each of May and June for stock monitoring purposes. The closure was initially designated to last for three years but has been repeatedly extended and remains in force. The level of effort of the monitoring fishery was increased in 2006.

## Natura 2000

To protect habitats several Natura 2000 areas have been defined. It is still under negotiation which fisheries will be prohibited in these areas exactly. It is likely that for each of these areas different rules will apply.

## Unilateral management

In addition to the EU-wide statutory regulations, some countries impose additional management schemes on their fleets. One example of this is the Scottish Conservation Credits scheme which encompasses technical regulation and temporary spatial closures in return for derogation from some EU effort controls. This scheme, and others are described in the stock sections to which they pertain.

### 2.3 Environmental considerations

WGNSSK welcomes the progress made to provide ecosystem overviews. These overviews give a good overview on environmental factors influencing the current development of fish and shellfish stocks. However, from these overviews it is still difficult to relate certain changes to observations made in the assessments. Therefore, the WG considers that although it is clear that the North Sea ecosystem is undergoing change and this will affect fish stocks, the causal mechanisms linking the environment with fish stock dynamics are in most cases not yet clearly-enough understood. However, for gadoids the choice of appropriate reference points takes now into account the current low productivity of the stocks although the exact causes of this low productivity are not fully understood. Fmsy is estimated based on shortened stock recruitment time series and the upper range of $\mathrm{F}_{\mathrm{msy}}$ is constraint by Fp 05 estimated for the current low recruitment period. To improve the situation, ICES may provide a database with all available environmental data and indicators from the various working and study groups to make them available to the scientific community. The longer the time series and the higher the contrast in these time series, the more likely that causal relationships can be identified.

Next to this WGNSSK made the following observations during the discussion on the ecosystem overviews:

1 ) The current low productivity of gadoids in the North Sea is not mentioned in the document. In general, under impact on commercial stocks an overview figure showing recruitment trends for the different guilds could provide valuable information.

2 ) The word crustaceans should be replaced with Nephrops in Figure 6.1.1.7. Only for Nephrops assessments are available and Nephrops constitutes only a small part of the crustacean biomass.

3 ) No flatfish are in the figure showing the North Sea food web. This is questionable for a flatfish dominated system .
4 ) The OSPAR table on threatened and declining species needs a review. Some of the species should not be mentioned any more (e.g., thornback ray, cod)
5 ) The ranking of the strength of interactions between pressure and state are pure qualitative (Figure 6.1.1.3). It is also unclear who has decided on the ranking. There are excellent expert elicitation methods available that could be applied to get an objective ranking based on the opinions of ICES experts.
6 ) The overviews need to be a living document. New knowledge needs to be incorporated when becoming available.

### 2.4 Human consumption fisheries

### 2.4.1 Data

Estimates of discarding rates provided by a number of countries through observer sampling programme were used in the assessments of various roundfish and flatfish as well as Nephrops FUs, to raise landings to catch (see also section 01 on Intercatch). During recent benchmarks discards could be included in the assessments of sole in 4, saithe in 4, 3.a and 6a as well as plaice in 7d. Discards could be also estimated for bycatch species (e.g., dab, flounder, lemon sole, witch, brill, turbot). Finally, catch advice could be given for all WGNSSK stocks.

In the EU, national sampling programs are defined and implemented as part of the Data Collection Framework (DCF). Other sampling programmes (e.g. industry selfsampling for discards and biological data) have been in place in recent years and the data are increasingly entering the assessment process in some instances (e.g., plaice in 4 , haddock). In general, some discarding occurred in most human-consumption fisheries until 2015. As TACs have become more restrictive for some species (e.g. cod), an increase in discarding of marketable fish (i.e. over minimum landing size) has been observed. In 2013, a landing obligation has been agreed between the EU Parliament and the Council of Ministers, as one of the most important aspects of the reform of the Common Fishery Policy (CFP), and this is going to have fundamental implications for the demersal fisheries and associated data collection program (see above).

For a number of years there had been indications that substantial under-reporting of roundfish and flatfish landings is likely to have occurred. It is suspected to have been particularly strong for cod until 2006, and catches were expected to be larger than the TAC. Since the middle of the 2000s, the WG had used an assessment method for North Sea cod (Section 14) which estimated unallocated removals, potentially due to to reporting problems, unrecorded discards, changes in natural mortality, or changes in survey catchability etc. In 2013, WGNSSK considered that the assumption of unallocated removals after 2006 could not be justified by any known factors (cf also ICES WKCOD 2011), and relaxed that assumption in the assessment.

Several research-vessel survey indices are available for most species, and were used both to calibrate population estimates from catch-at-age analyses, and in exploratory analyses based on survey data only. Commercial CPUE series were available for a number of fleets and stocks, but for various reasons few of them could be used for assessment purposes (although they are presented and discussed). The use of commercial CPUE indices has been phased out where possible and only the saithe and sole in 7 d assessment still relies on a commercial index.

Bycatches in the industrial fisheries were significant in the past for haddock, whiting and saithe, but these have reduced considerably in recent years.

### 2.4.2 Summary of stock status

The main impression in 2016 is that fishing mortality has been reduced significantly for many North Sea stocks of roundfish and flatfish compared to the beginning of the century. All fish stocks with agreed biomass reference points are above $\mathrm{B}_{\mathrm{lim}}$, and only the SSBs of haddock in 4, 6.a and 3.a. 20 and sole in 7d are below MSY Btrigger at the beginning of 2016 (the update assessment for cod, given in Annex 04, shows it is now above MSY Btrigger). The assessment of haddock was initially rejected by WGNSSK and the stock status was re-evaluated in autumn after an Interbenchmark. Several North Sea stocks are exploited around or below Fmsy levels. Exceptions are cod in 4, 3.a. 20 and 7 d , haddock in 4, 6.a and 3.a.20, whiting in 4 and $7 . \mathrm{d}$ and sole in 7d. An important feature is that recruitment still remains poor compared to historic average levels for most gadoids. For whiting, the most recent recruitments are estimated to be higher but still do not reach historically observed high recruitment levels; however, the decreasing trend in SSB has stopped and the whiting stock has increased in recent years. The stock status of saithe has been revised during the 2016 benchmark process. According to the latest assessment, the saithe stock has fluctuated without trend above MSY Btrigger since 1997. The southern North Sea and eastern channel areas are currently experiencing a steep increase of the plaice stock, with SSB values higher than ever observed in the assessment time series. Following new survey information during the summer, advice was re-opened for cod in 4, 3.a. 20 and 7.d, saithe in 4,6 and 3.a, and Nephrops FUs 6, 7 and 8, and this are given in Annex 04.

WGNSSK is also responsible for the assessment of several flatfish species that are mainly by catch in demersal fisheries. For all of these stocks, catch advice was provided in 2015 for the first time. The trends in abundance/biomass indices is decreasing for some of these stocks and increasing for others. In 2016 it was only necessary to determine whether the perception of the stock has changed compared to 2015; because perceptions have not changed compared to 2015, no reopening of the advice was needed.

The stock status of Nephrops stocks differs between functional units but globally a decrease compared to earlier years is observed for Nephrops in the North Sea, although a strong increase has been observed in the latest UWTV survey for FU7, which may be related to a strong recruitment event.

The summary of stock status is as follows:
1 ) Nephrops: Abundance globally decreased compared to former years in the North Sea. The abundance of Nephrops in FU 6 remains low, but there have been recent increases in FU7 and 8, while FU9 has been stable in recent years. The abundance of Nephrops in 3.a is stable and at a high level. The 2015 harvest rates were in accordance with FMSY in FU 7, FU 9 and in 3.a. In contrast, harvest rates in FU 6 exceeded FMSY in 2015. For FU 8 the harvest rate was very close to Fmsy. Because the TAC is set for the whole North Sea and not
at a functional unit level, this contributes to Fmsy reference points being exceeded in some FUs. The FUs 5, 32, 33 and 34 are data limited. For these FUs, densities have to be assumed, along with other variables (e.g. discard rates, mean weights) and/or TV-survey estimates are uncertain. Given these assumptions, current harvest rates seem not to be problematic, apart from those for FU6.
2 ) Norway Pout in 4 and 3.a: The stock size is highly variable from year to year, due to recruitment variability and a short life span. Stock size has increased and is above $B_{p a}$ in 2016. Fishing mortality has been below the long-term average $\mathrm{F}(0.45)$ since 1995 . Recruitment in 2014 and 2016 are high, while recruitment in 2015 is around average.

3 ) Cod in 4, 3.a. 20 and 7d: Fishing mortality has been declining since 2000 and is estimated to be above Fmsy. Spawning-stock biomass has increased from the historical low in 2006 and is just above MSY Btrigger. Recruitment since 1998 remains poor.
4 ) Haddock in 4, 6.a and 3.a.20: Fishing mortality is above FMSY and spawningstock biomass has fallen below MSY Btrigger. Recruitment since 2000 has been characterized by a low average level with occasional larger year classes, the size of which is diminishing. The 2014 recruitment estimate is higher than recent poor recruitment years, but is still below the long-term average.
5 ) Whiting in 4 and 7d: Spawning-stock biomass has fluctuated around MSY $B_{\text {trigger. }}$ Fishing mortality has been above Fmsy throughout the time-series. Recruitment has been low since 2003, with recruitment in 2014 and 2015 above previous years.
6 ) Saithe in 3.a, 4 and 6: Recruitment has fluctuated over time and has generally been below the long-term average since 2008. Fishing mortality has been below FmsY since 2013. Spawning-stock biomass has fluctuated without trend, remaining above MSY Btrigger since 1997.
7 ) Plaice in 4 and 3.a.20: The combined North Sea and Skagerrak stock is well above MSY $B_{\text {trigger, }}$ has increased in the past ten years, and has been at a record high for the last 5 years. Recruitment has been around the long-term average since the mid-90s. In recent years, fishing mortality has been estimated around Fmsy.
8 ) Sole in 4: The spawning stock biomass has increased since 2007 and has been estimated at above MSY Btrigger since 2012. Fishing mortality has declined since 1997 and is estimated at FMSY in 2015. Recruitment has fluctuated with- $^{2}$ out trend since the early 1990s
9 ) Plaice in 7d: Fishing mortality has declined since the mid-1990s and has been below Fmsy since 2009. Spawning-stock biomass has increased since 2008 and has been above MSY $B_{\text {trigger }}$ since 2009. Recruitment has been high since 2009.

10 ) Sole in 7d: The spawning-stock biomass has fluctuated without trend and is predicted to drop below MSY $\mathrm{B}_{\text {trigger }}$ in 2016. Fishing mortality has always been above F msy and increased over the years 2012-2015. Recruitment has been fluctuating without trend, and was in 2012-2014 among the lowest of the time series, which has resulted in the decrease in recent SSB.
11 ) Category 3-6 stocks: In 2016 a new advice has been produced for Pollack in 4 and 3.a as well as grey gurnard in 4, 7d and 3.a. Pollack is a category 5 stock. However, information available suggests that the stock is at a low
level compared to historic times. The time series of mature biomass index of grey gurnard from the IBTS-Q1 survey showed a strong increase from the beginning of 90's, and has since fluctuated at a high level. Advice on other category 3 stocks will be updated next year. New survey information has not changed the perception of the stocks.

### 2.5 Industrial fisheries

The Norway Pout assessment was benchmarked in 2012 through an inter-benchmark protocol (IBPNPOUT), resulting in changes in biological parameters (growth, maturity and natural mortality), and again in 2016 (WKPOUT) during which the assessment model was changed, but the general perception of the stock hasn't changed substantially. Fishery has fluctuated considerably in recent years with full or partial closures in 2005, 2006, 2007 and 2011. The stock is largely driven by natural process, particularly recruitment. Stock size has increased and is above $B_{p a}$ in 2016. Fishing mortality has been below the long-term average F (0.45) since 1995. Recruitment in 2014 and 2016 are high, while recruitment in 2015 is around average.

### 2.6 Input from The ICES - FAO Working Group on Fishing Technology \& Fish Behaviour (WGFTFB)

The WGFTFB provided fishery development information specific to the various assessment Expert Groups every second year, based on annual questionnaires to a number of FTFB members (ICES 2012, WGFTFB). Unfortunately in 2015 and 2016 no information was provided. The information specific to WGNSSK in 2013 was the following:
This report outlines a number of technical issues relating to fishing technology that may impact on fishing mortality and more general ecological impacts. This includes information recent changes in commercial fleet behaviour that may influence commercial CPUE estimates; identification of recent technological advances (creep); ecosystem effects; and the development of new fisheries in the North Sea, Skagerrak and Kattegat.

It should be noted that the information contained in this report does not cover fully all fleets engaged in North Sea fisheries; information was obtained from Scotland, France, Netherlands, Belgium, Sweden, Denmark and Norway.

## Fleet dynamics

- It is now apparent that within the Netherlands, driven primarily by the cost of fuel, there is huge demand to use the pulse trawl and the number of vessels applying to fish under the $5 \%$ derogation far exceeds the number of licences available. Vessels not using the pulse trawl in the Netherlands are finding it increasingly difficult to get financial support from banks on economical (high fuel prices making beam trawling uneconomic) and ecological grounds (beam trawls are portrayed negatively). A total of 42 licences have been given out by the Dutch Ministry for pulse trawling. The majority of these licences are for flatfish beam trawls although 2-3 vessels are involved in trials to test the Belgium "HOVERCRAN" system. (Netherlands: Implications: Switch from beam trawling to pulse trawling).
- Measures to reduce fuel costs in the Dutch fleet have been continuing since 2008 from $35 \%$ (now $50 \mathrm{M} €$ ) to $25 \%$ in 2009 due to adaptations in gear and operation. Reports show that the use of the SumWing can save up to 300 tonnes of fuel per year per boat ( $\mathrm{Loa}=40 \mathrm{~m}$ ), and with the pulse trawl up to 800-1000 tonnes annually. Up to 78 Dutch beam trawlers now used the SumWing with 28 of these using the Pulse/Wing trawl. A further 12 Dutch registered vessels in the UK, 2 in Germany and 5 in Belgium are also using the Sum Wing with 3 of these using the dual Pulse/Wing. (Netherlands: Implications: Switch to more fuel efficient gear).
- The use of the SumWing, Pulse Trawl and Pulse/Wing beam trawls are reported to have resulted in a shift in grounds in ICES Area 4, and also add fishing time, due to faster hauling speed. This is not well documented but may result in increased a change in effort patters and increased fishing time (Netherlands: Implications: Improved fuel efficiency).
- A recent analysis carried out in the Netherlands has measured the relative changes in catching efficiency between the pulse trawl and the standard beam trawl in numbers caught per hour. This analysis has shown a reduction in catching efficiency for plaice of around $25-30 \%$; for sole $20-25 \%$ with a reduction in discards of $55-60 \%$ and a reduction in benthos of $35-40-\%$. (Netherlands: Implications: Changes in catching efficiency).
- There are a smaller number of vessels in Belgium, Germany and the UK (Dutch owned) using the pulse trawl. However, the derogation only allows it to be used in the south of the North Sea and this has hindered the uptake of the gear in Belgium as they have only a limited sole quota in this area. (Belgium: Implications: Lower uptake of pulse trawl).
- The shift to Danish/Scottish seining has continued in a number of countries. This is driven by high fuel prices: There are now around 16 Dutch vessels operating in the North and the South North Sea, while there are around 12 French vessels now converted to Danish seining. (France: Implications: Shift from trawling to Seining).
- At least two Dutch beam trawlers have been converted into dual purpose vessels with the capability of fishing with twin-rigs and beam trawls. The idea is to increase the fishing opportunities for these vessels and allow them to switch between fisheries at different time of the year. (Netherlands: Implications: Dual purpose fishing vessels).
- In Norway there has not been any removal of fishing effort through decommissioning but the number of vessels has reduced due to movement of quota from newer and bigger vessels. In particular the medium and larger sized seine net fleets are being renewed. Altogether 3-4 new large whitefish trawlers, 4-5 pursers, 7-9 large seine netters and a large number of coastal vessels have entered the fleet over the period 2010-2012. While not measured it is likely that capacity and effort have increased. (Norway: Implications: Increases in fishing effort).


## Technology Creep

- As reported the use of the SumWing and Pulse trawl are widespread now in the Netherlands. This is driven by high fuel costs with reported savings ranging from $10-50 \%$ from a combination of the gear and lower towing speeds. (Netherlands: Implications: Improved fuel efficiency).
- Since 2010 one large Belgium beam trawler is now using the energy-efficient SumWing as a replacement for standard beam trawl gear. In addition some 25 other beam trawlers are using the SumWing seasonally. Additional there is growing interest in using the electric pulse trawl amongst the Belgium beam trawl fleet allthough none of them are using as yet as they are awaiting authorisation to do so from the EU. These changes are driven by high fuel prices. (Belgium: Implications: Changes in gear type).
- There have been continued efforts by many countries to reduce fuel costs through the use of more energy efficient gears. Modifications tested include
hydrodynamic/low impact trawl doors, dynex warps, low drag twines and reductions in the size of trawls used. In addition there have been changes in fishing operations through slower steaming to and from grounds, fishing closer to home ports and other fuel saving measures. The actual impacts on fishing effort are difficult to predict but this is likely to be a continuing trend. However, as a result of one project in France (EFFICHALUT) an energy efficient trawl has been developed and is now being used by 15 whitefish trawls. The anticipated annual savings from these 15 vessels is estimated at $€ 800000$ per year. (Multiple countries: Implications: Improved energy efficiency)
- IMR continues to cooperate with a commercial partner (Scantrol AS) on development of a camera-based system to identify and measure individual fish inside a trawl. Preliminary results indicate lengths estimated using the camera system are within $5-10 \%$ of manually measured lengths. The system has the potential to be a useful tool to verify size and species stratification by depth during acoustic surveys. A new, more compact, system capable of operating at up to 2000 m depth is being readied for field trials in May and software development to further automate image analysis is underway. (Norway: Implications: New acoustic survey tool).


## Technical Conservation Measures

- The Dutch government closed the Botany Gut in December 2011 and 2011 for three months to all trawling to protect cod. The Nephrops fishermen have looked for an exemption from this closure on the basis of low cod impact although observer data suggest catches of 2-5/hour Pilot studies will be carried out in 2012 to test selective gears to reduce cod catches. (Netherlands: Implications: Reduction of cod catches).
- At the Dutch Shipyard, Maaskant are developing an on board separator system to sort out debris and benthic organisms before they affect the catch. The objective of this system is to improve the survival rate of discarded flatfish. Trials will be carried out during 2012 and 2013. (Netherlands: Implications: Improved survival of discarded fish).
- In Sweden the use of rigid sorting grids has continued to increase in the Skagerrak and Kattegat. In Skagerrak the use of sorting grid has increased from $50 \% 2009$ to $53 \% 2010$ of total TR2 effort. The TR2 effort is by far the most important gear category in both Skagerrak and Kattegat constituting $80-90 \%$ of total effort. Almost $100 \%$ have opted to use this device due primarily to national legislation allocating $50 \%$ of the total Nephrops quota to grid vessels (Sweden: Implications: Widespread use of sorting grid).
- Extensive testing has been carried out in the North Sea by the UK (Scotland) with a new design of trawl called the Flip-flap netting grid (FFG) which has been developed by a Scottish netmaker for reducing cod catches in the Nephrops fishery. The results show a large and significant decrease in the number of the three main whitefish species retained by the FFG gear. The reductions by weight of cod, haddock and whiting were 73,67 and $82 \%$ respectively. There are indications that will be introduced as a regulated gear across the Scottish fleet during 2012. (UK (Scotland): Implications: Widespread use of selective gear)
- Trials with sorting grids in the Norway pout fishery were completed in 2011, with further tests of flexible grid designs. In 2011 grids with a 40 mm bar
spacing were made mandatory for Norwegian vessels on the basis of these and earlier trials for blue whiting and Norway pout in the Barents Sea. From mid-2013 all vessels with have to use in these small mesh fisheries in the Norwegian Economic Zone in the North Sea. (Norway: Implications: Mandatory use of sorting grids).
- Experiments from the North Sea carried out by Denmark during 2011have indicated that $50 \%$ of Nephrops above MLS are lost through the codend in nominal 120 mm codends used in the North Sea indicated a general increase in mesh size is not applicable in this fishery. (Denmark: Implications: Loss of marke Table catch)
- In July 2011 Denmark introduced a mandatory regulation requiring the use of the SELTRA trawl in the demersal fisheries in the Kattegat. This selective device comprises a 180 mm square mesh panel or 270 mm diamond mesh panel contained in a 4 panel section laced 4 m from the codend. This on the basis of trials with this device which showed very good reductions in cod catches. The L50 for the 180 mm panel is around 64 cm . (Denmark: Implications: Mandatory introduction of a selective gear)


## Ecosystem effects

- The large scale uptake of the pulse trawl has resulted in reduced ecosystem impacts on benthos and also a reduction in discards. A large scale monitoring programme is currently being undertaken to fully assess these reductions (Netherlands: Implications: Reduced ecosystem impacts in the southern North Sea).
- Several Dutch beam trawlers are testing a new T-Line concept using pins instead of chains to chase fish out of the seabed. The initial trials carried out in December 2011 were not particular successful as the pins had a tendency to break off and catches of sole were low but a new version will be tested later in 2012. The objective of this design is to reduce fuel and impact on the seabed. (Netherlands: Implications: Lower ecosystem impact of beam trawls).
- Uptake of the Swedish cod and haddock quotas in the Skagerrak have been high in the first two quarters of 2012 (between 55-70\%) raising fears that the fisheries will be closed early in the year leading to discarding (Sweden: Implications: Potential discard problem).


## Development of New Fisheries

- Two Dutch vessels have converted to potting for crab due to high fuel prices and low returns from other fishing methods. No other details are reported (Netherlands: Implications: Testing of new fisheries).
- The fishery for greater weever (Trachinus draco) in the Kattegat seen during 2010 and 2011 has continued. This fishery has developed as a consequence of low catches of Nephrops and cod during the first quarter of the year. The weever is also one of few species that are without limiting quotas and few regulations attached to it in the Kattegat. (Sweden: Implications: development of new fishery).


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## 3 Norway lobster (Nephrops spp.) in Division 3.a (Skagerrak and Kattegat)

### 3.1 General comments relating to all Nephrops stocks (3.a and 4)

### 3.1.1 Introduction

Nephrops stocks have previously been identified by WGNEPH on the basis of their population distribution and characteristics, and established as separate Functional Units. The Functional Units (FU) are defined by the groupings of ICES statistical rectangles given in Table 3.1.1 and illustrated in Figure 3.1.1. The statistical rectangles making up each FU encompass the distribution of mud sediment on which Nephrops live. There are two FUs in Division 3.a and nine FUs in Subarea 4. At the 2010 WG, it was noted that a significant and increasing proportion of Nephrops landings were being taken outside the previously defined FUs in Subarea 4. This has led to the introduction of a new FU (FU 34) covering the Devil's Hole. Additional catches of Nephrops are also taken from smaller, isolated pockets of mud distributed throughout the ICES divisions (e.g. off the east coast of Scotland at Arbroath). Management of Nephrops currently operates at the ICES Subarea/Division level.

Functional Units were previously aggregated by WGNEPH into a series of nominal Management Areas (MA) intended to provide a pragmatic solution for more localised management. In 2008 the Working Group agreed that this process had served no useful purpose and should be discontinued.

MSY estimation for Nephrops stocks is complicated by the absence of an age-based analytical assessment. The process for determining suitable $\mathrm{F}_{\text {msy }}$ proxies for Nephrops stocks can be found in Section 3.3.4.

### 3.1.2 A new approach for data poor Nephrops stocks

The WKLIFE considered the following Nephrops stocks: FU 5 (Botney Gut - Silver Pit), 10 (Noup), 32 (Norwegian Deep), and 33 (Off Horns Reef). All four stocks were considered to belong to category 6 (data-limited stocks) including stocks for which only landings data are available. The working group agrees with this classification. WKLIFE considered the available data for these stocks. An L50 value (Length at $50 \%$ maturity) exists for Nephrops in FU 5, otherwise there is no information on growth parameters or maturity. The newly established functional unit 34 (Devil's Hole) is also a category 6 data poor stock.
According to WKLIFE, SPR and FsPR reference points have been identified as proxies for SSBmsy and Fmsy respectively. These reference points could be used to inform risk assessment approaches applied to category 6 and 7 stocks and can be calculated on the basis of life-history information and knowledge of selection patterns. Life-history traits (LHTs) should be compiled by stock experts in the relevant assessment working groups. LHTs are available from a number of sources including Fish-Base, literature not (yet) accounted in FishBase, grey literature, and recent estimates based on DCF data collection.

In 2014 the working group introduced a different approach to previous years in order to provide an estimated guidance of the biomass in FUs 5, 10, 32, 33, and 34 and consider different harvest rates. Using FU area (calculated from information on the extension of suitable habitat and/or extent of Nephrops fisheries), mean discard percentage from all
years of data, and mean weight in catches, tables of harvest rates were calculated for each of the five data poor functional units, using a range of landings ( 100 t to maximum landings observed for each stock) and densities ( $0.05-0.8$ animals $\mathrm{m}^{-2}$ ). The density range comes from the North Sea/Skagerrak stocks for which UWTV surveys exist. For each data poor FU, the mean and maximum of the landings time series is marked in the table. Harvest rates larger than $10 \%$ are marked red. For each stock the most likely densities are considered based on information from neighbouring FUs.

This approach enables the working group to consider the sustainability of historic landings as well as to present guidance to landings within safe biological limits.

The presentation of specific data and assessments relating to the Divisions 3.a and 4 FUs can be found in the WGNSSK report sections 3 and 4, respectively.

### 3.2 Nephrops in Division 3.a

### 3.2.1 General

At present there are two functional units in Division 3.a: Skagerrak (FU 3) and Kattegat (FU 4). This separation was based on observed differences between Skagerrak and Kattegat regarding Nephrops size composition in catches in the 1980s and 1990s. However, the distribution of Nephrops is almost continuous from southern Kattegat into Skagerrak, and the exchange of pelagic larvae between the southern and northern areas is very likely. With the longer data series now available, it seems the differences in size composition between the two areas are more likely to be random or caused by factors from fishing operations. The assessment is therefore conducted on Nephrops in 3.a as one stock.

## Ecosystem aspects

Nephrops live in burrows in suitable muddy sediments and is characterised by being omnivorous and emerge out of the burrows to feed. It can, however, also sustain itself as a suspension feeder in the burrows (Loo et al., 1993). This ability may contribute to maintaining a high production of this species in 3.a, due to increased organic production. Nephrops have recently been found to have a high prevalence of plastics which may have implications for the health of the stock (Murry and Cowie, 2011).

Severe depletion in oxygen content in the water can force the animals out of their burrows, thus temporarily increasing the trawl catchability of this species during such environmental changes (Bagge et al. 1979). An especially severe case was observed in the end of the 1980s in the southern part of 3.a in late summer, where unusually high catch rates of Nephrops were observed. The increasing amount of dead specimens in the catches led to the conclusion of severe oxygen deficiency in especially the southern part of 3.a (Kattegat) in late 1988 (Bagge et al., 1990).

No information is available on the extent to which larval mixing occurs between Nephrops stocks, but the similarity in stock indicator trends between FU 3 and 4 for both Denmark and Sweden indicates that recruitment has been similar in both areas. These observations suggest they may be related to environmental influences.

## ICES Advice

The most recent advice for Nephrops in 3. a was given in 2015. ICES concluded that:
'Stock size is considered to be stable. The estimated harvest ratios suggest that the fishing mortality ( F ) for this stock is currently below FMSY.'

## Management for FU 3 and FU 4

The TAC for Nephrops in ICES area 3.a was increased from 5318 t in 2015 to 11001 t in 2015. The large increase in quota was due to the fact that the EU shifted from providing landings advice to providing catch advice. The minimum conservation reference size (previously referred to as minimum landings size) for Nephrops in area 3.a was reduced in 2016 from 40 to 32 mm carapace length. The historically large MLS led to a high discard ratios (discards/ discards+landings), and during $201541 \%$ of the catch (in numbers) in 3.a consisted of undersized individuals (Figure 3.2.1.1). The reduction in MLS is expected to reduce the proportion of the catch discarded considerably. Furthermore, it is expected that ongoing experimental work on improving gear selectivity will further reduce the amount discarded. A discard ban was implemented in EU waters from the $1^{\text {st }}$ January 2015. The discard ban became applicable to Nephrops from the $1^{\text {st }}$ January 2016, however an exemption for high survivability was introduced for one year. New technical measures have also been agreed upon and have been implemented since the $1^{\text {st }}$ February 2013.

Swedish gear regulations since 2004 imply that it is mandatory to use a 35 mm species selective grid together with an 8 m full square-mesh codend of 70 mm and extension piece when trawling for Nephrops in Swedish national waters. Additionally, the Danish gear regulations since 2011 imply a mandatory use of either the grid or the use of the SELTRA trawl which compromise a 90 mm cod end with either a square-mesh panel ( 180 mm in the Kattegat and 140 mm in the Skagerrak) or 270 mm diamond mesh panel. In Article 11 in the cod recovery plan, member states may apply for unlimited number of days when using the species selective grid trawl.

### 3.2.2 Data available from Skagerrak (FU3) and Kattegat (FU4)

## Landings

Division 3.a includes FU 3 and 4, which are assessed together. Total Nephrops landings by FU and country are shown in Table 3.2.1.1 and Table 3.2.1.2.

FU 3 is primarily exploited by Denmark, Sweden and Norway. Denmark and Sweden dominate this fishery, with $58 \%$ and $38 \%$ by weight of the landings in 2015, respectively. Landings by the Swedish creel fishery represented 13-18 \% of the total Swedish Nephrops landings from the Skagerrak in the period 1991 to 2002. Since 2002 creel catches have been steadily increasing and have in 2009 to 2014 accounted for more than $30 \%$, and in 2015 for more than $40 \%$ of Swedish Skagerrak landings (Table 3.2.2.1). In the early 1980s, total Nephrops landings from the Skagerrak increased from around 1000 t to just over 2670 t . Since then they have been fluctuating around a mean of 2500 t (Figure 3.2.2.1).

Both Denmark and Sweden have Nephrops directed fisheries in the FU 4 (Kattegat). In 2015, Denmark accounted for about 73 \% of total landings in FU4, while Sweden took 27 \% (Table 3.2.2.5). Minor landings have been taken by Germany (<1\%).

After a decline in the observed landings in 1994, total Nephrops landings from the Kattegat increased again until 1998 and have fluctuated around 1500 t . However, since 2006 the landings have increased and were in 2010 the highest on record over the 50 year period (Figure 3.2.2.4). Since 2010, landings show a decreasing trend.

## Length compositions

For the Skagerrak, size distributions of both the landings and discards are available from both Denmark and Sweden for 1991-2015. In the beginning of the time series, the Swedish data can be considered as being the most complete, since sampling took place regularly throughout the time period and usually covered the whole year. Trends in mean size in catch and landings for Skagerrak are shown in Figure 3.2.2.2 and Table 3.2.2.4. Mean sizes for landings are fluctuating without trend. Mean size for undersized show an increasing trend since 2005.

For Kattegat, size distributions of both the landings and discards are available from Sweden for 1990-2015, and from Denmark for 1992-2015. The at-sea-sampling intensity has generally increased since 1999. The Danish sampling intensity was low in 2007 and 2008, but was normalized in 2009 to 2015. Information on mean size is shown in Figure 3.2.2.5 and Table 3.2.2.8. Notice, that except for small mean sizes from 1993 to 1996 all categories have since been fluctuating without trend.

In earlier years, the Swedish discard samples were obtained by agreement with selected fishermen, and this might have tempted fishermen to bias the samples. However, the reliability of the catch samplings was cross-checked by special discard sampling projects in both the Skagerrak and the Kattegat. In recent years, the Swedish Nephrops sampling has been carried out by onboard observers in both Skagerrak and Kattegat. In 1991, a biological sampling programme of the Danish Nephrops fishery was started on board fishing vessels in order to also cover the discards in this fishery. Due to its high cost and the lack of manpower, Danish sampling intensity in the early years was in general not satisfactory, and seasonal variations were not often adequately covered. The Norwegian Nephrops fishery is small and has not been sampled.

## Natural mortality, maturity at age and other biological parameters

In previous analytical assessments (when Length Cohort Analyses were performed, see e.g. WGNEPH, 2003), natural mortality was assumed to be 0.3 for males of all ages and in all years. Natural mortality was assumed to be 0.3 for immature females, and 0.2 for mature females. Discard survival was assumed to be 0.25 for both males and females (after Gueguen \& Charuau, 1975, Redant \& Polet, 1994, and Wileman et al. 1999).

Growth parameters are as follows:
Males: $\quad \mathrm{L} \infty=73 \mathrm{~mm} \mathrm{CL}, \mathrm{k}=0.138$.
Immature females: $\quad \mathrm{L} \infty=73 \mathrm{~mm} \mathrm{CL}, \mathrm{k}=0.138$.
Mature females: $\quad \mathrm{L} \infty=65 \mathrm{~mm}$ CL, $\mathrm{k}=0.10$, Size at $50 \%$ maturity $=29 \mathrm{~mm}$ CL.
Growth parameters for males were taken from Ulmestrand and Eggert (2001) and female growth parameters have been assumed to be similar to those of Scottish Nephrops stocks.

Data on size at maturity for males and females were presented at the ICES Workshop on Nephrops Stocks in January 2006 (ICES WKNEPH, 2006).

## Catch and effort data - FU3

Effort data for the Swedish fleet are available from logbooks for 1978-2015 (Figure 3.2.2.1 and Table 3.2.2.2). During the period 1998 to 2005, twin trawlers shifted to targeting both fish and Nephrops, which resulted in a decreasing trend in LPUE during this period (Table 3.2.2.2). Since 2005, LPUE for twin trawls has increased. The LPUE for single trawls has shown and increasing trend throughout the entire time series. The long term trend
in LPUEs is similar in the Swedish and Danish fisheries (Figure 3.2.2.1). Total Swedish trawl effort shows a decreasing trend since 1992 and has been fluctuating without trend since 2003. From 2007 onwards, total Swedish trawl effort has been estimated from LPUEs from the single trawl with a grid (targeting only Nephrops).
Danish effort figures for the Skagerrak (Table 3.2.2.3 and Figure 3.2.2.1) were estimated from logbook data. For the whole period, it is assumed that effort is exerted mainly by vessels using twin trawls. The overall trend in effort for the Danish fleet is similar to that in the Swedish fishery. After having been at a relatively low level in 1994-98, effort increased again in the next four years, followed by a decrease to a relatively low level in 2007 to 2015. Also the trend in LPUE is similar to that in the Swedish single trawl fishery, however with a much more marked increase in the Danish LPUE for 2007 and 2008. This high LPUE level is likely to be a consequence of the national (Danish) management system introduced in 2007.

It has not been possible to explicitly incorporate 'technological creeping' in a further evaluation of the Danish effort data. However, since 2000 the Danish logbook data have been analysed in various ways to elucidate the effect of factors likely to influence the effort/LPUE, e.g. vessel size (Figure 3.2.2.3).

## Catch and effort data - FU4

Swedish total effort has been relatively stable over the period 1978-90. Effort increased from 1990 to 1993, followed by a decrease to 1996. During the last 20 years effort has remained relatively stable, except for 2007 and 2008 where effort increased (Figures 3.2.2.4 and Table 3.2.2.6). Figures for total Danish effort are based on logbook records since 1987. Danish effort increased from 1995 to 2001, decreased from 2002 to 2007 and has been fluctuating without trend since (Figure 3.2.2.4 and Table 3.2.2.7).

Since 2000 the Danish logbook data have been standardised to account for changes in fishing power due to changes in the physical characters of the Nephrops fleet. The data have been analysed in various ways to elucidate the effect of factors likely to influence the effort/LPUE, e.g. vessel size (Figure 3.2.2.6).

### 3.2.3 Combined assessment (FU $3 \& 4$ )

## Reviews of last year's assessment

"No major issues. It was noted that it would be useful to show confidence intervals around the UWTV estimates. The LPUE considerations were moved to additional considerations."

### 3.2.3.1 TV survey in 3.a

In 2008 and 2009, an exploratory UWTV survey was carried out by Denmark. In 2010, the TV survey was expanded covering the main Nephrops grounds in the western part of Skagerrak (subarea 1) and Northern part of Kattegat (subarea 2). Since 2011, the TV survey has been carried out in collaboration between Denmark and Sweden and covers the main Nephrops fishing grounds in 3.a (subarea 1-6). In 2014, subarea 1 was extended to the west (subarea 7; Figure 3.2.3.2). However, important parts of the assumed distributional range of Nephrops were still not covered in 2015. The survey is still developing and improved spatial coverage is expected to be raised during the 2016 benchmark. Figure 3.2.3.4 presents the distribution of stations with valid density estimates from 2008 to 2015. A similar survey design has been applied for both national surveys: a fixed grid with random stratified stations.

In order to estimate the total population numbers, the density estimates have to be raised from the survey areas to total area of the population distribution. VMS information is currently the best available proxy to estimate the Nephrops stock distribution in 3.a. VMS data from the Swedish and Danish fishery from 2010 were used (Figure 3.2.3.3) and are described in more detail in ICES (2011). The area estimates for each subarea are defined in Table 3.2.3.1. Burrow counting and identification follows the standard protocols defined by WGNEPS (ICES, 2013).

## Abundance indices from UWTV surveys

The number of valid stations conducted in the UWTV survey in 3.a divided into subareas Figure 3.2.3.2 is shown in Table 3.2.3.1 and Figure 3.2.3.4.

In WKNEPH (2009) a number of bias sources were highlighted relating to the "counted" density from the TV surveys. These bias sources are not easily estimated and are largely based on expert opinion. For the Nephrops stock in 3.a it is assumed that the largest source of perceived bias is the "edge effect", due to the relative large sizes of the burrow systems. The cumulative biases result in a correction factor to take the raw counts to absolute densities. The correction factor for 3 .a was set to be 1.1, meaning that the raw TV survey is likely to overestimate Nephrops abundance by $10 \%$. TV survey results are presented as absolute values (i.e. the bias already taken into account)

| FU | Area | Edge <br> EFFECT | Detection RATE | SPECIES IDENTIFICATION | Occupanc Y | Cumulative BIAS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 and 4 | Skagerrak <br> and <br> Kattegat <br> (3.a) | 1.3 | 0.75 | 1.05 | 1 | 1.1 |

### 3.2.3.2 2015 Assessment.

The assessment of the state of the Nephrops stock in 3.a is based on the UWTV survey from 2015. Additional used information was trends in total combined (Denmark and Sweden) LPUE, and discards (numbers) as a proxy for recruitment during the period 1990-2015.

Combined relative effort declined slightly over the period 1990 to 2015 (Figure 3.2.4.1) while combined relative LPUE shows an increasing trend and is at a high level in the recent 7 years (Figure 3.2.4.2). This high level may be attributed to the change in the Danish management system (Individual Transferable Quotas) in 2007. Technical creep, changes in targeting behaviour, stock size and catchability may also be responsible for some of this increase. High LPUEs attributable to sudden changes in catchability (caused by e.g. poor oxygen conditions) are known to occur but are generally of short duration.

Since the abundance of small Nephrops (typically discards of specimens below minimum landing size) may also be regarded as an index of recruitment, they can be used to further explain the current developments in the stock. The large amounts of discards in the periods 1993-95 and 1999-2000 reflect strong recruitment during these years (Figure 3.2.4.3). The high levels of discards in 1993-95 are believed to have significantly contributed to the high LPUE in 1998-99. The high amount of discards observed in 2007, 2008 and 2009 would then indicate high recruitment in these years, as could the low amount of discards in 2014 and 2015 indicate a low recruitment. The discards in 2015 is the lowest since 1991 and may be due to a very low recruitment and/or an increase in gear size selectivity.

## MSY consideration (TV-survey)

There are no precautionary reference points defined for Nephrops. Under the ICES MSY framework, exploitation rates which are likely to generate high long-term yields (and low probability of stock overfishing) have been explored and proposed for Division 3.a. Owing to the way Nephrops are assessed, it is not possible to estimate FMSY directly and hence proxies for FMSY are determined. WGNSSK (2010) developed a framework for proposing Fmsy proxies for the various Nephrops stocks based upon their biological and historical characteristics, and is described in section 1 of that report. Three candidates for $\mathrm{F}_{\text {MSY }}$ are $\mathrm{F}_{0.1}, \mathrm{~F}_{35 \% \text { Spr }}$ and $\mathrm{F}_{\mathrm{max}}$. There may be strong differences in relative exploitation rates between the sexes in many stocks. To account for this, values for each of the candidates have been determined for males, females and the two sexes combined. An appropriate Fmsy candidate has been selected according to the perception of stock resilience, factors affecting recruitment, population density, knowledge of biological parameters and the nature of the fishery (relative exploitation of the sexes and historical harvest rate vs stock status).

A decision-making framework based on the table below was used in the selection of preliminary stock-specific Fmš proxies (ICES, 2010a). These proxies may be modified following further data exploration and analysis. The combined sex Fmsy proxy should be considered appropriate if the resulting percentage of virgin spawner-per-recruit for males or females does not fall below $20 \%$. When this does happen a more conservative sex-specific Fmsy proxy should be picked instead of the combined proxy.

|  |  | Burrow density (average burrows m-2) |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Low | Medium | High |
|  |  | $<0.3$ | 0.3-0.8 | >0.8 |
| Observed harvest rate or landings compared to stock status | > Fmax | F35\%SPR | Fmax | Fmax |
|  | Fmax - F0.1 | F0.1 | F35\%SPR | Fmax |
|  | < F0.1 | F0.1 | F0.1 | $\begin{aligned} & \text { F35\%SP } \\ & \text { R } \end{aligned}$ |
|  | Unknown | F0.1 | F35\%SPR | $\begin{aligned} & \text { F35\%SP } \\ & \text { R } \end{aligned}$ |
| Stock size estimates | Variable | F0.1 | F0.1 | F35\% |
|  | Stable | F0.1 | F35\%SPR | Fmax |
| Knowledge of biological parameters | Poor | F0.1 | F0.1 | $\begin{aligned} & \mathrm{F} 35 \% \mathrm{SP} \\ & \mathrm{R} \end{aligned}$ |
|  | Good | F35\%SPR | F35\%SPR | Fmax |
| Fishery history | Stable spatially and temporally | F35\%SPR | F35\%SPR | Fmax |
|  | Sporadic | F0.1 | F0.1 | $\begin{aligned} & \text { F35\%SP } \\ & \text { R } \end{aligned}$ |
|  | Developing | F0.1 | F35\%SPR | $\begin{aligned} & \text { F35\%SP } \\ & \text { R } \end{aligned}$ |

The absolute burrow density in Division 3.a is medium $\left(0.3-0.8 / \mathrm{m}^{2}\right)$, the observed harvest rate is below $\mathrm{F}_{0.1}$ and historically the fishery is stable both spatially and temporally. This means that $\mathrm{F}_{0.1}$ may be selected as a proxy for $\mathrm{F}_{\text {MSY. }}$ As the MLS has been decreased in 2016 and this stock will be benchmarked during 2016, it is recommended to use Fmax as a proxy for FMSY as in last years. For 2017 this corresponds to a TAC of 13099 tonnes if a landing obligation is applied. Under a landings obligation it may well be necessary to recalculate a harvest rate associated with FMSY as total catches would be subjected to $100 \%$ mortality (current discard survival is estimated to be $25 \%$ ).

A mismatch between mesh size in trawl fisheries and minimum landing size (carapace length 40 mm , which is higher than North Sea FUs) has historically resulted in a high discard proportion for this stock. However, since 1st January 2016 the MCRS/MLS was lowered from 40 to 32 mm carapace length for EU countries. This is expected to reduce the proportion of the catch discarded considerably. Norway still apply 40 mm MCRS and a Norwegian discard ban was implemented in the Skagerrak since 1st of January 2015.

To simulate the effect of a decreased MCRS on the proportion of discards, the average (2013-2015) total sampled length distribution (graph left below) was first used to estimate fisher's selection when sorting the catch at a MCRS of 40 mm carapace length (red line in middle graph below). This selection ogive was then shifted down to 32 mm MCRS (assuming that fishers selection is equally effective at the new MCRS) in order to predict the new composition of landings and discards (see graph right below). The following mean weight in discards and landings, discard proportion and dead discard rate was used in this year's assessment.


Recent Swedish discard survival experiments indicate that the trawl discard survival may be higher (around $50 \%$ ) compared to the $25 \%$ currently used in the assessment. This has caused a possibility to continue discard Nephrops $<$ MCRS (high survivability exemption) during 2016. Effects of discard survival estimates will be discussed during the coming benchmark meeting in 2016.

Harvest rate as proxy for FMš for 3.a from length cohort analysis 2011 (2008-2010):

|  |  | MALE |  |
| :--- | :--- | :--- | :--- |
| Fmax | $6.8 \%$ | $10.0 \%$ | Combined |
| F0.1 | $4.9 \%$ | $7.6 \%$ | $7.9 \%$ |
| F35\%SpR | $8.1 \%$ | $12.9 \%$ | $5.6 \%$ |

The harvest rates ((landings + dead discards)/total stock biomass) equivalent to $\mathrm{F}_{\text {msy }}$ proxies are based on yield-per-recruit analyses from length cohort analyses. These analyses utilise average length frequency data taken over the 3 year period (2008-2010). All Fmsy proxy harvest rate values are considered preliminary and may be modified following further data exploration and analysis.

Norway lobster in Division 3.a. The catch options.

| VARIABLE | VALUE | SOURCE | Notes |
| :--- | :--- | :--- | :--- |
| Abundance in TV <br> assessment | 3857 <br> million | ICES 2016a | UWTV 2015 |
| Mean weight in <br> landings* | 46.2 g | ICES 2016a | Average 2013-2015 |
| Mean weight in <br> discards* | 20.5 g | ICES 2016a | Average 2013-2015 |
| Discard proportion* | $12.5 \%$ | ICES 2016a | Average (proportion by number) <br> 2013-2015 |
| Discard survival rate | $25 \%$ | ICES 2016a | Proportion by number. Only applies <br> in scenarios where discarding <br> allowed. |
| Dead discard rate* | $9.7 \%$ | ICES 2016a | Average 2013-2015 (proportion by <br> number), Only applies in scenarios <br> where discarding allowed |

* Simulated that MCRS was 32 mm carapace length during 2013-2015.

Landings obligation

| Basis | Total CATCH* | WANTED CATCH | UnWANTED CATCH | HARVEST RATE* |
| :--- | :--- | :--- | :--- | :--- |
|  | L+D | L | D | for L + DD |
| F2015 | 3399 | 3196 | 203 | $2.1 \%$ |
| Fcurrent (2013-2015) | 6019 | 5660 | 359 | $3.6 \%$ |
| F0.1 = FmsyLower | 9284 | 8731 | 553 | $5.6 \%$ |
| MSY Approach | 13099 | 12318 | 781 | $7.9 \%$ |
| F35SpR | 17410 | 16372 | 1038 | $10.5 \%$ |

## Weights in tonnes

* as calculated for dead removals


## Discarding allowed

| BASIS | Total CATCH* | Dead removals | Landings | $\begin{aligned} & \text { DeAD } \\ & \text { DISCARDS* } \end{aligned}$ | Surviving DISCARDS* | Harvest RATE** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | L+DD+SD | L+DD | L | DD | SD | for L+DD |
| MSY approach | 13521 | 13319 | 12715 | 604 | 201 | 7.9\% |

Weights in tonnes.
*Total discard ratio is assumed to be $12.5 \%$ of the catches (by number, average of last three years, 20132015), MCRS is changed to 32 mm carapace length, discard survival (SD) is assumed to be $\mathbf{2 5 \%}$ (WKNEPH; ICES, 2009).
** as calculated for dead removals
A summary of the results from the TV survey 2015 is presented in Table 3.2.3.1. The estimated abundance index was 0.393 resulting in a total abundance of 3857 million individuals. Total removals (landings + dead discards) were estimated to 79 million individuals resulting in a harvest rate of $2.0 \%$.

## Conclusions drawn from the indicator analyses

The combined logbook recorded effort has decreased by $50 \%$ since 2002 and is currently at a low level while LPUE shows an increasing trend and is at a long term high level in recent years (Figures 3.2.4.1 and 3.2.4.2). Mean sizes are fluctuating without trend. There are no signs of overexploitation in 3.a.

The conclusion from this indicator based assessment is that the stock is exploited sustainably.

### 3.2.4 Biological reference points

No biological reference points are used for this stock.

### 3.2.5 Quality of the assessment

The length and sex composition of the landings data is considered to be well sampled. Discard sampling in this fishery has been conducted on a quarterly basis for Danish and Swedish Nephrops trawlers since 1990, and is considered to represent the fishery adequately.

The UWTV survey 2015 was conducted in all 7 defined subdivisions in 3.a. A correction factor of 1.1 was used. A total weighted mean density was estimated based on density estimates from each subdivision and weighted by the size of each subdivision. The estimated $\mathrm{F}_{\mathrm{msy}}$ proxies for this stock provide a relatively low harvest rate which may be a result of the high discards ratios ( $31 \%$ in weight) which occur due to the high minimum landing size $(40 \mathrm{~mm})$. These removals do not increase the yield from the stock.

The Danish LPUE data used as indicators for stock development have been standardised regarding engine size. However, LPUE is also influenced by changes in catchability due to sudden changes in the environmental conditions or/and changes in selectivity, gear efficiency or a change in targeting behaviour due to the cod management plan in 3.a. Also the changes in management systems (indicated by the broken red line in Figure 3.2.4.2), which occurred in 2007 in Denmark, caused a general increase in LPUE. In 3.a, fluctuations in catches of small Nephrops are used as indicators of recruitment (Figure 3.2.4.3).

### 3.2.6 Status of the Stock

The Nephrops stock in Division 3.a was assessed with an UWTV survey for the fifth year (2011-2015; new subarea 7 only in 2014-2015) and the time series of UWTV estimates is still insufficient to draw conclusions regarding stock trajectory (Figure 3.3.6.1).

The 2015 harvest rate was estimated to be relatively low ( $2.0 \%$ from UWTV survey) implying the stock appears to be exploited sustainably.

The analysis of commercial LPUE and effort data indicate that LPUE shows an increasing trend while effort shows a decreasing trend and the WG concludes that current levels of exploitation appear to be sustainable.

### 3.2.7 Division 3.a Nephrops Management Considerations

The observed trends in effort, LPUE and discards are similar for FU 3 and FU 4. Our present knowledge on the biological characteristics of the Nephrops stocks in these two areas does not indicate obvious differences, and therefore the two FUs are treated as one single 'stock' in the assessment.
The UWTV- survey in 3.a suggests that the harvest rate of the stock is relatively low and the stock is exploited at a sustainable level.
The combined logbook recorded effort has decreased since 2002 and is currently the lowest level in the time series while LPUE has increased and is at a relatively high level in the last ten years (Figures 3.2.4.1 and 3.2.4.2). Mean sizes are fluctuating without trend (Figures 3.2.2.2 and 3.2.2.5). There are no signs of overexploitation in 3.a.

Given the apparent stability of the stock, the WG concludes that current levels of exploitation appear to be sustainable.

The WG encourages the work on size selectivity in Nephrops trawls to reduce the large amount of discarded undersized Nephrops in 3.a.

## Mixed fishery aspects

Cod and sole are significant by-catch species in these fisheries in 3.a, and even if data on catches, including discards, of the by-catch gradually become available, they have not yet been used in the management. The WG has for many years recommended the use of species selective grids in the fisheries targeting Nephrops as legislated for Swedish national waters. New technical measures (Swedish grid and SELTRA trawl) have recently been agreed upon for the Nephrops directed fishery and have been implemented since the 1st February 2013. The European Union and Norway have also agreed that a discard ban will be implemented in EU waters from the 1st January 2015. The discard ban will be applicable to Nephrops from the 1st January 2016 but preliminary results indicating high discard survival has resulted in an exception of landing obligation for Nephrops in 3.a during 2016.

Table 3.1.1. Definition of Nephrops Functional Units in 3.a and 4 in terms of ICES statistical rectangles.

| FU No. | Name | ICES AREA | Statistical rectangles |
| :---: | :---: | :---: | :---: |
| 3 | Skagerrak | 3.aN | 47G0; 46F9-G1; 45F8-G1; 44F7-G0; 43F8-F9 |
| 4 | Kattegat | 3.aS | 44G1; 42-43 G0-G2; 41G1-G2 |
| 5 | Botney Gut - Silver Pit | 4b,c | 36-37 F1-F4; 35F2-F3 |
| 6 | Farn Deeps | 4b | 38-40 E8-E9; 37E9 |
| 7 | Fladen Ground | 4a | 44-49 E9-F1; 45-46E8 |
| 8 | Firth of Forth | 4b | 40-41E7; 41E6 |
| 9 | Moray Firth | 4a | 44-45 E6-E7; 44E8 |
| 10 | Noup | 4a | 47E6 |
| 32 | Norwegian Deep | 4a | 44-52 F2-F6; 43F5-F7 |
| 33 | Off Horn Reef | 4b | 39-41F5; 39-41F6 |
| 34 | Devil's Hole | 4b | 41-43 F0-F1 |

Table 3.2.1.1. Division 3.a: Total Nephrops landings (tonnes) by Functional Unit, 1981-2015.

| Year | FU 3 | FU 4 | Total |
| :---: | :---: | :---: | :---: |
| 1981 | 992 | 1728 | 2720 |
| 1982 | 1470 | 1828 | 3298 |
| 1983 | 2205 | 1472 | 3677 |
| 1984 | 2675 | 2036 | 4711 |
| 1985 | 2191 | 1798 | 3989 |
| 1986 | 2018 | 1807 | 3825 |
| 1987 | 2441 | 1605 | 4046 |
| 1988 | 2363 | 1364 | 3727 |
| 1989 | 2564 | 1313 | 3877 |
| 1990 | 2866 | 1475 | 4341 |
| 1991 | 2924 | 1304 | 4228 |
| 1992 | 1893 | 1012 | 2905 |
| 1993 | 2288 | 924 | 3212 |
| 1994 | 1981 | 893 | 2874 |
| 1995 | 2429 | 998 | 3427 |
| 1996 | 2695 | 1285 | 3980 |
| 1997 | 2612 | 1594 | 4206 |
| 1998 | 3248 | 1808 | 5056 |
| 1999 | 3194 | 1755 | 4949 |
| 2000 | 2894 | 1816 | 4710 |
| 2001 | 2282 | 1774 | 4056 |
| 2002 | 2977 | 1471 | 4448 |
| 2003 | 2126 | 1641 | 3767 |
| 2004 | 2312 | 1653 | 3965 |
| 2005 | 2546 | 1488 | 4034 |
| 2006 | 2392 | 1280 | 3672 |
| 2007 | 2771 | 1741 | 4512 |
| 2008 | 2851 | 2025 | 4876 |
| 2009 | 3004 | 1842 | 4846 |
| 2010 | 2938 | 2185 | 5123 |
| 2011 | 2511 | 1475 | 3986 |
| 2012 | 2536 | 1893 | 4429 |
| 2013 | 2147 | 1613 | 3760 |
| 2014 | 2856 | 1294 | 4150 |
| 2015 | 2123 | 1228 | 3350 |

Table 3.2.1.2. Division 3.a: Total Nephrops landings (tonnes) by country, 1991-2015.

| Year | Denmark | Norway | Sweden | Germany | Total landings | Total Disc. | Total Catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 2824 | 185 | 1219 |  | 4228 | 5183 | 9411 |
| 1992 | 2052 | 104 | 749 |  | 2905 | 2523 | 5428 |
| 1993 | 2250 | 103 | 859 |  | 3212 | 8493 | 11705 |
| 1994 | 2049 | 62 | 763 |  | 2874 | 6450 | 9324 |
| 1995 | 2419 | 90 | 918 |  | 3427 | 4464 | 7891 |
| 1996 | 2844 | 102 | 1034 |  | 3980 | 2148 | 6128 |
| 1997 | 2959 | 117 | 1130 |  | 4206 | 3469 | 7675 |
| 1998 | 3541 | 184 | 1319 | 12 | 5056 | 1944 | 7000 |
| 1999 | 3486 | 214 | 1243 | 6 | 4949 | 4108 | 9057 |
| 2000 | 3325 | 181 | 1197 | 7 | 4710 | 5664 | 10374 |
| 2001 | 2880 | 138 | 1037 | 1 | 4056 | 3767 | 7823 |
| 2002 | 3293 | 116 | 1032 | 7 | 4448 | 4311 | 8760 |
| 2003 | 2757 | 99 | 898 | 13 | 3767 | 2208 | 5975 |
| 2004 | 2955 | 95 | 903 | 12 | 3965 | 2532 | 6497 |
| 2005 | 2901 | 83 | 1048 | 2 | 4034 | 3014 | 7048 |
| 2006 | 2432 | 91 | 1143 | 6 | 3672 | 2926 | 6598 |
| 2007 | 2887 | 145 | 1467 | 13 | 4512 | 6524 | 11036 |
| 2008 | 3174 | 158 | 1509 | 19 | 4860 | 4746 | 9606 |
| 2009 | 3372 | 128 | 1331 | 15 | 4846 | 6129 | 10975 |
| 2010 | 3721 | 124 | 1249 | 29 | 5123 | 3548 | 8671 |
| 2011 | 2937 | 87 | 945 | 17 | 3986 | 2847 | 6833 |
| 2012 | 2970 | 104 | 1355 | 0 | 4429 | 4771 | 9200 |
| 2013 | 2550 | 73 | 1134 | 3 | 3760 | 4010 | 7770 |
| 2014 | 2785 | 88 | 1269 | 7 | 4150 | 1854 | 6004 |
| 2015 | 2121 | 91 | 1138 | 0 | 3350 | 1038 | 4389 |

Table 3.2.2.1. Nephrops in Skagerrak (FU 3): Landings (tonnes) by country, 1991-2015.

| YeAR | Denmark | Norway |  |  | Sweden |  |  | Germany | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Trawl | Creel | Sub-total | Trawl | Creel | Sub-total |  |  |
| 1991 | 1639 | 185 | 0 | 185 | 949 | 151 | 1100 | 0 | 2924 |
| 1992 | 1151 | 104 | 0 | 104 | 524 | 114 | 638 | 0 | 1893 |
| 1993 | 1485 | 101 | 2 | 103 | 577 | 123 | 700 | 0 | 2288 |
| 1994 | 1298 | 62 | 0 | 62 | 531 | 90 | 621 | 0 | 1981 |
| 1995 | 1569 | 90 | 0 | 90 | 659 | 111 | 770 | 0 | 2429 |
| 1996 | 1772 | 102 | 0 | 102 | 708 | 113 | 821 | 0 | 2695 |
| 1997 | 1687 | 117 | 0 | 117 | 690 | 118 | 808 | 0 | 2612 |
| 1998 | 2055 | 184 | 0 | 184 | 864 | 145 | 1009 | 0 | 3248 |
| 1999 | 2070 | 214 | 0 | 214 | 793 | 117 | 910 | 0 | 3194 |
| 2000 | 1877 | 181 | 0 | 181 | 689 | 147 | 836 | 0 | 2894 |
| 2001 | 1416 | 125 | 13 | 138 | 594 | 134 | 728 | 0 | 2282 |
| 2002 | 2053 | 99 | 17 | 116 | 658 | 150 | 808 | 0 | 2977 |
| 2003 | 1421 | 90 | 9 | 99 | 471 | 135 | 606 | 0 | 2126 |
| 2004 | 1595 | 85 | 10 | 95 | 449 | 173 | 622 | 0 | 2312 |
| 2005 | 1727 | 71 | 12 | 83 | 538 | 198 | 736 | 0 | 2546 |
| 2006 | 1516 | 80 | 11 | 91 | 583 | 201 | 784 | 0 | 2391 |
| 2007 | 1664 | 127 | 18 | 145 | 709 | 253 | 962 | 0 | 2771 |
| 2008 | 1745 | 124 | 34 | 158 | 675 | 273 | 948 | 0 | 2851 |
| 2009 | 2012 | 101 | 27 | 128 | 605 | 260 | 864 | 0 | 3004 |
| 2010 | 1981 | 105 | 20 | 125 | 563 | 266 | 829 | 4 | 2938 |
| 2011 | 1801 | 74 | 12 | 87 | 432 | 188 | 621 | 2 | 2510 |
| 2012 | 1516 | 80 | 24 | 104 | 592 | 324 | 916 | 0 | 2536 |
| 2013 | 1309 | 57 | 16 | 73 | 484 | 279 | 763 | 0 | 2146 |
| 2014 | 1868 | 68 | 20 | 88 | 594 | 305 | 899 | 0 | 2856 |
| 2015 | 1226 | 66 | 25 | 91 | 479 | 327 | 806 | 0 | 2123 |

Table 3.2.2.2. Nephrops Skagerrak (FU 3): Catches and landings (tonnes), effort ('000 hours trawling), CPUE and LPUE (kg/hour trawling) of Swedish Nephrops trawlers, 1991-2015. (*Include only Nephrops trawls with grid and square mesh codend).

| Single trawl |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Catches | Landings | Effort | CPUE | LPUE |
| 1991 | 676 | 401 | 71.4 | 9.5 | 5.6 |
| 1992 | 360 | 231 | 73.7 | 4.9 | 3.1 |
| 1993 | 614 | 279 | 72.6 | 8.4 | 3.8 |
| 1994 | 441 | 246 | 60.1 | 7.3 | 4.1 |
| 1995 | 501 | 336 | 60.8 | 7.8 | 5.2 |
| 1996 | 754 | 488 | 51.1 | 14.8 | 9.6 |
| 1997 | 643 | 437 | 44.4 | 14.4 | 9.8 |
| 1998 | 794 | 557 | 49.7 | 16.0 | 11.2 |
| 1999 | 605 | 386 | 34.5 | 17.5 | 9.3 |
| 2000 | 486 | 329 | 32.7 | 14.9 | 10.9 |
| 2001 | 446 | 236 | 26.2 | 17.0 | 10.4 |
| 2002 | 503 | 301 | 29.4 | 17.1 | 8.8 |
| 2003 | 310 | 254 | 21.5 | 13.9 | 11.4 |
| 2004* | 474 | 257 | 20.1 | 23.6 | 13.4 |
| 2005* | 760 | 339 | 29.7 | 25.6 | 12.7 |
| 2006* | 839 | 401 | 37.5 | 22.4 | 12.2 |
| 2007* | 894 | 314 | 24.1 | 37.0 | 13.0 |
| 2008* | 605 | 264 | 20.0 | 30.3 | 13.2 |
| 2009* | 482 | 285 | 19.6 | 24.5 | 14.5 |
| 2010* | 476 | 286 | 20.7 | 23.0 | 13.8 |
| 2011* | 334 | 198 | 16.8 | 19.9 | 11.8 |
| 2012* | 542 | 238 | 16.0 | 33.8 | 14.9 |
| 2013* | 251 | 137 | 11.3 | 22.2 | 12.1 |
| 2014* | 240 | 157 | 11.0 | 21.7 | 14.2 |
| 2015* | 187 | 133 | 9.5 | 19.6 | 14.0 |
| Twin trawl |  |  |  |  |  |
| Year | Catches | Landings | Effort | CPUE | LPUE |
| 1991 | 740 | 439 | 39.5 | 18.7 | 11.1 |
| 1992 | 370 | 238 | 34.1 | 10.9 | 7.0 |
| 1993 | 568 | 258 | 35.9 | 15.8 | 7.2 |
| 1994 | 444 | 248 | 34.1 | 13.1 | 7.3 |
| 1995 | 403 | 270 | 32.9 | 12.2 | 8.2 |
| 1996 | 187 | 121 | 13.0 | 14.4 | 9.3 |
| 1997 | 219 | 149 | 17.5 | 12.5 | 8.5 |
| 1998 | 254 | 178 | 16.7 | 15.2 | 10.6 |
| 1999 | 382 | 244 | 27.6 | 13.8 | 8.8 |
| 2000 | 349 | 237 | 31.3 | 11.1 | 10.1 |
| 2001 | 470 | 249 | 33.7 | 14.0 | 7.4 |
| 2002 | 392 | 244 | 33.3 | 11.8 | 7.1 |
| 2003 | 168 | 138 | 22.5 | 7.5 | 6.1 |
| 2004 | 217 | 118 | 21.7 | 10.0 | 5.4 |
| 2005 | 263 | 117 | 22.1 | 11.9 | 5.3 |
| 2006 | 253 | 121 | 19.6 | 12.9 | 6.2 |
| 2007* | 248 | 87 | 5.4 | 45.6 | 16.0 |
| 2008* | 139 | 61 | 3.4 | 41.3 | 18.0 |
| 2009* | 211 | 125 | 7.1 | 29.5 | 17.5 |
| 2010* | 165 | 99 | 5.9 | 27.8 | 16.7 |
| 2011* | 202 | 120 | 7.7 | 26.3 | 15.6 |
| 2012* | 544 | 239 | 12.9 | 42.2 | 18.6 |
| 2013* | 423 | 231 | 13.8 | 30.7 | 16.8 |
| 2014* | 484 | 316 | 16.0 | 30.3 | 19.8 |
| 2015* | 328 | 234 | 11.3 | 28.9 | 20.6 |

Table 3.2.2.3. Nephrops Skagerrak (FU 3): Logbook recorded effort (kW days, Days at sea, and fishing days) and LPUE ( $\mathrm{kg} /$ day) for bottom trawlers catching Nephrops with codend mesh sizes of 70 mm or above, and estimated total effort by Danish trawlers, 1991-2015.

| Year | KW DAYS | DAYS AT SEA | FISHING DAYS | LPUE |
| :---: | :---: | :---: | :---: | :---: |
| 1991 | 5501223 | 21043 | 18762 | 87 |
| 1992 | 4043742 | 16125 | 13970 | 82 |
| 1993 | 3728965 | 13698 | 11958 | 124 |
| 1994 | 3276355 | 12324 | 10778 | 120 |
| 1995 | 3024232 | 12070 | 10448 | 150 |
| 1996 | 3020019 | 11871 | 10385 | 171 |
| 1997 | 3053570 | 11950 | 10509 | 161 |
| 1998 | 3353072 | 12131 | 10899 | 189 |
| 1999 | 3967797 | 13767 | 12376 | 167 |
| 2000 | 4371006 | 14849 | 13307 | 141 |
| 2001 | 3970228 | 13337 | 11579 | 122 |
| 2002 | 4693962 | 16575 | 14197 | 145 |
| 2003 | 3476385 | 11589 | 10333 | 138 |
| 2004 | 3871974 | 13149 | 11694 | 136 |
| 2005 | 3757466 | 12560 | 11166 | 155 |
| 2006 | 3296744 | 10825 | 9725 | 156 |
| 2007 | 2424063 | 8026 | 7294 | 228 |
| 2008 | 2332056 | 8016 | 7300 | 239 |
| 2009 | 2549895 | 8814 | 8058 | 250 |
| 2010 | 2668904 | 9027 | 8338 | 238 |
| 2011 | 2666680 | 9767 | 8912 | 202 |
| 2012 | 2183682 | 8330 | 7507 | 202 |
| 2013 | 1738286 | 6770 | 6332 | 207 |
| 2014 | 2094860 | 8060 | 7653 | 244 |
| 2015 | 1592065 | 6337 | 5923 | 207 |

Table 3.2.2.4. Skagerrak (FU 3): Mean sizes (mm CL) of male and female Nephrops in catches of Danish and Swedish combined, 1991-2015.

| Year | Catches |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Undersized |  | Full sized |  | All |  |
|  | Males | Females | Males | Females | Males | Females |
| 1991 | 30.2 | 30.9 | 41.2 | 42.7 | 30.9 | 29.8 |
| 1992 | 33.3 | 32.3 | 43.3 | 44.7 | 33.3 | 32.2 |
| 1993 | 33.0 | 31.5 | 42.0 | 43.6 | 33.0 | 31.5 |
| 1994 | 31.7 | 29.6 | 41.7 | 43.6 | 31.7 | 29.6 |
| 1995 | 30.0 | 28.5 | 41.6 | 41.3 | 32.9 | 29.8 |
| 1996 | 33.2 | 31.9 | 42.9 | 44.0 | 37.6 | 37.0 |
| 1997 | 35.8 | 34.5 | 44.6 | 44.1 | 39.8 | 39.1 |
| 1998 | 34.8 | 34.4 | 46.1 | 43.9 | 40.7 | 37.3 |
| 1999 | 34.6 | 33.9 | 44.9 | 43.8 | 39.3 | 36.1 |
| 2000 | 30.6 | 30.5 | 45.6 | 45.0 | 32.5 | 34.1 |
| 2001 | 33.6 | 33.6 | 45.5 | 43.6 | 37.3 | 36.4 |
| 2002 | 33.9 | 33.7 | 44.0 | 42.5 | 37.2 | 37.3 |
| 2003 | 33.5 | 32.6 | 43.2 | 43.4 | 38.0 | 36.7 |
| 2004 | 34.3 | 33.4 | 44.6 | 45.2 | 38.7 | 36.6 |
| 2005 | 33.5 | 32.4 | 43.7 | 43.0 | 36.4 | 35.3 |
| 2006 | 33.2 | 32.9 | 44.7 | 42.7 | 37.1 | 36.1 |
| 2007 | 32.6 | 31.9 | 44.4 | 42.4 | 34.9 | 33.5 |
| 2008 | 33.6 | 32.3 | 44.0 | 42.7 | 36.5 | 34.5 |
| 2009 | 35.0 | 33.8 | 45.3 | 42.8 | 39.8 | 35.9 |
| 2010 | 34.2 | 33.8 | 46.2 | 44.8 | 38.9 | 36.6 |
| 2011 | 33.8 | 33.1 | 44.5 | 43.3 | 38.4 | 36.5 |
| 2012 | 34.8 | 34.1 | 44.2 | 42.5 | 38.2 | 36.2 |
| 2013 | 35.1 | 34.8 | 45.0 | 42.9 | 38.6 | 36.9 |
| 2014 | 35.7 | 35.3 | 45.5 | 43.7 | 41.7 | 39.1 |
| 2015 | 35.5 | 36.2 | 47.2 | 44.1 | 43.6 | 41.1 |

Table 3.2.2.5. Nephrops Kattegat (FU 4): Landings (tonnes) by country, 1991-2015.

| Year | Denmark | Sweden |  | Sub-total | Germany | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Trawl | Creel |  |  |  |
| 1991 | 1185 | 119 | 0 | 119 | 0 | 1304 |
| 1992 | 901 | 111 | 0 | 111 | 0 | 1012 |
| 1993 | 765 | 159 | 0 | 159 | 0 | 924 |
| 1994 | 751 | 142 | 0 | 142 | 0 | 893 |
| 1995 | 850 | 148 | 0 | 148 | 0 | 998 |
| 1996 | 1072 | 213 | 0 | 213 | 0 | 1285 |
| 1997 | 1272 | 319 | 3 | 322 | 0 | 1594 |
| 1998 | 1486 | 306 | 4 | 310 | 12 | 1808 |
| 1999 | 1416 | 329 | 4 | 333 | 6 | 1755 |
| 2000 | 1448 | 357 | 4 | 361 | 7 | 1816 |
| 2001 | 1464 | 304 | 6 | 309 | 1 | 1774 |
| 2002 | 1240 | 219 | 5 | 224 | 7 | 1471 |
| 2003 | 1336 | 287 | 5 | 292 | 13 | 1641 |
| 2004 | 1360 | 270 | 11 | 281 | 12 | 1653 |
| 2005 | 1175 | 303 | 8 | 311 | 2 | 1488 |
| 2006 | 916 | 347 | 11 | 358 | 6 | 1280 |
| 2007 | 1223 | 491 | 15 | 505 | 13 | 1741 |
| 2008 | 1429 | 561 | 16 | 577 | 19 | 2025 |
| 2009 | 1360 | 450 | 16 | 467 | 15 | 1842 |
| 2010 | 1740 | 403 | 17 | 420 | 25 | 2185 |
| 2011 | 1136 | 308 | 16 | 324 | 15 | 1475 |
| 2012 | 1454 | 406 | 33 | 439 | 0 | 1893 |
| 2013 | 1241 | 341 | 27 | 368 | 3 | 1612 |
| 2014 | 917 | 335 | 34 | 369 | 7 | 1294 |
| 2015 | 895 | 301 | 31 | 333 | 0 | 1228 |

Table 3.2.2.6. Kattegat (FU 4): Catches and landings (tonnes), effort ('000 hours trawling), CPUE and LPUE (kg/hour trawling) of Swedish Nephrops trawlers, 1991-2015 (*Include only Nephrops trawls with grid and square mesh codend).

| SINGLE TRAWL |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Catches | Landings | Effort | CPUE | LPUE |
| 1991 | 66 | 39 | 10.3 | 6.4 | 3.7 |
| 1992 | 44 | 28 | 11.6 | 3.8 | 2.4 |
| 1993 | 128 | 58 | 14.9 | 8.6 | 3.9 |
| 1994 | 95 | 53 | 16.2 | 5.7 | 3.2 |
| 1995 | 79 | 53 | 9.6 | 7.8 | 5.5 |
| 1996 | 207 | 134 | 13.7 | 15.1 | 9.8 |
| 1997 | 269 | 183 | 18.0 | 15.0 | 10.2 |
| 1998 | 181 | 127 | 13.1 | 13.8 | 9.7 |
| 1999 | 146 | 93 | 8.1 | 17.9 | 11.4 |
| 2000 | 114 | 77 | 8.5 | 13.4 | 9.1 |
| 2001 | 117 | 62 | 7.6 | 15.4 | 8.2 |
| 2002 | 42 | 25 | 3.7 | 11.2 | 6.7 |
| 2003 | 49 | 40 | 4.6 | 10.7 | 8.7 |
| 2004 | 70 | 44 | 4.3 | 16.2 | 10.1 |
| 2005 | 147 | 100 | 12.3 | 11.9 | 8.1 |
| 2006 | 234 | 154 | 15.1 | 15.5 | 10.2 |
| 2007* | 107 | 51 | 4.1 | 25.7 | 12.3 |
| 2008* | 121 | 57 | 4.4 | 27.6 | 13.0 |
| 2009* | 157 | 81 | 5.1 | 30.9 | 16.1 |
| 2010* | 181 | 102 | 7.6 | 23.8 | 13.4 |
| 2011* | 75 | 45 | 3.8 | 20.0 | 12.0 |
| 2012* | 80 | 45 | 3.4 | 23.5 | 13.3 |
| 2013* | 44 | 26 | 2.3 | 19.5 | 11.6 |
| 2014* | 35 | 25 | 2.2 | 15.8 | 11.6 |
| 2015 | 43 | 29 | 2.6 | 16.6 | 11.0 |
| Twin trawl |  |  |  |  |  |
| Year | Catches | Landings | Effort | CPUE | LPUE |
| 1991 | 93 | 55 | 8.8 | 10.6 | 6.2 |
| 1992 | 101 | 65 | 14.2 | 7.1 | 4.6 |
| 1993 | 187 | 85 | 17.8 | 10.6 | 4.8 |
| 1994 | 138 | 77 | 14.2 | 9.7 | 5.4 |
| 1995 | 125 | 84 | 11.0 | 12.2 | 7.7 |
| 1996 | 97 | 63 | 7.5 | 13.0 | 8.4 |
| 1997 | 183 | 124 | 12.7 | 14.3 | 9.7 |
| 1998 | 215 | 151 | 15.0 | 14.4 | 10.1 |
| 1999 | 306 | 195 | 20.1 | 15.2 | 9.7 |
| 2000 | 330 | 224 | 24.5 | 13.5 | 9.1 |
| 2001 | 353 | 187 | 25.1 | 14.1 | 7.4 |
| 2002 | 256 | 153 | 23.2 | 11.0 | 6.6 |
| 2003 | 222 | 181 | 24.8 | 8.9 | 7.3 |
| 2004 | 253 | 158 | 16.5 | 15.4 | 9.6 |
| 2005 | 198 | 135 | 15.3 | 12.9 | 8.8 |
| 2006 | 183 | 121 | 12.7 | 14.4 | 9.5 |
| 2007* | 112 | 54 | 3.6 | 30.9 | 14.8 |
| 2008* | 164 | 78 | 4.8 | 34.1 | 16.1 |
| 2009* | 309 | 161 | 11.0 | 28.2 | 14.6 |
| 2010* | 297 | 167 | 9.2 | 32.2 | 18.1 |
| 2011* | 266 | 159 | 9.7 | 27.3 | 16.3 |
| 2012* | 406 | 231 | 12.4 | 32.8 | 18.6 |
| 2013* | 354 | 210 | 15.0 | 23.7 | 14.0 |
| 2014* | 282 | 206 | 14.4 | 19.6 | 14.4 |
| 2015 | 262 | 173 | 11.3 | 23.2 | 15.4 |

Table 3.2.2.7. Nephrops Kattegat (FU 4): Logbook recorded effort (kW days, Days at sea, and fishing days) and LPUE ( $\mathrm{kg} /$ day) for bottom trawlers catching Nephrops with codend mesh sizes of 70 mm or above, and estimated total effort by Danish trawlers, 1991-2015.

| Year | KW DAYS | Days at sea | Fishing days | LPUE |
| :---: | :---: | :---: | :---: | :---: |
| 1991 | 4223351 | 23040 | 16770 | 71 |
| 1992 | 3689413 | 20184 | 14240 | 63 |
| 1993 | 2827025 | 15392 | 10598 | 72 |
| 1994 | 2480847 | 13989 | 10985 | 68 |
| 1995 | 2330909 | 13023 | 10028 | 85 |
| 1996 | 2707363 | 14856 | 11688 | 92 |
| 1997 | 2807943 | 14389 | 11558 | 110 |
| 1998 | 2957280 | 15264 | 12380 | 120 |
| 1999 | 3417242 | 16734 | 13536 | 105 |
| 2000 | 3642120 | 18307 | 14661 | 99 |
| 2001 | 3826693 | 18764 | 15294 | 96 |
| 2002 | 3258819 | 16568 | 13325 | 93 |
| 2003 | 3173969 | 15345 | 12507 | 107 |
| 2004 | 2929407 | 14229 | 11289 | 120 |
| 2005 | 2452852 | 11814 | 9337 | 126 |
| 2006 | 2147461 | 10431 | 8467 | 108 |
| 2007 | 2022910 | 9883 | 7897 | 155 |
| 2008 | 2148132 | 10538 | 8469 | 169 |
| 2009 | 2219200 | 11120 | 8726 | 156 |
| 2010 | 2438736 | 12055 | 9707 | 179 |
| 2011 | 2009409 | 10286 | 8099 | 140 |
| 2012 | 2292229 | 11800 | 9661 | 150 |
| 2013 | 2221959 | 11669 | 9226 | 135 |
| 2014 | 1908170 | 10393 | 7865 | 117 |
| 2015 | 1847763 | 10094 | 7704 | 116 |

Table 3.2.2.8. Nephrops Kattegat (FU 4): Mean sizes (mm CL) of male and female Nephrops in discards, landings and catches, 1991-2015. Since 2005 based on combined Danish and Swedish data.

|  | CATCHES |  |  |  |  |  |  |  |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :---: | :---: |
|  | Piscards | Landings |  |  |  |  |  | All |
|  | Males | Females | Males | Females | Males | Females |  |  |
| 1991 | 30.7 | 31.1 | 42.4 | 42.5 | 32.5 | 32.9 |  |  |
| 1992 | 33.0 | 30.3 | 44.4 | 43.2 | 36.7 | 34.9 |  |  |
| 1993 | 30.5 | 29.3 | 42.3 | 43.1 | 31.3 | 30.1 |  |  |
| 1994 | 29.7 | 28.3 | 40.8 | 40.2 | 31.2 | 28.9 |  |  |
| 1995 | 30.8 | 30.5 | 42.4 | 42.0 | 33.7 | 33.2 |  |  |
| 1996 | 32.7 | 31.3 | 42.0 | 44.0 | 36.7 | 37.3 |  |  |
| 1997 | 33.6 | 33.2 | 45.0 | 44.5 | 37.1 | 35.0 |  |  |
| 1998 | 34.2 | 33.2 | 45.6 | 44.1 | 41.3 | 36.8 |  |  |
| 1999 | 32.9 | 33.8 | 45.3 | 40.9 | 37.8 | 34.9 |  |  |
| 2000 | 35.1 | 35.2 | 45.7 | 42.1 | 40.4 | 36.9 |  |  |
| 2001 | 32.2 | 33.0 | 44.1 | 41.9 | 35.9 | 36.5 |  |  |
| 2002 | 34.4 | 33.3 | 44.4 | 43.8 | 37.2 | 36.2 |  |  |
| 2003 | 33.0 | 33.2 | 43.5 | 42.2 | 37.1 | 36.0 |  |  |
| 2004 | 34.7 | 34.2 | 45.1 | 43.2 | 39.9 | 37.5 |  |  |
| 2005 | 33.5 | 33.9 | 45.8 | 43.1 | 38.7 | 38.7 |  |  |
| 2006 | 33.2 | 33.6 | 45.1 | 42.8 | 37.9 | 37.4 |  |  |
| 2007 | 33.9 | 33.2 | 44.8 | 43.5 | 37.2 | 35.5 |  |  |
| 2008 | 32.6 | 32.4 | 44.0 | 43.9 | 37.5 | 35.9 |  |  |
| 2009 | 33.8 | 33.1 | 44.7 | 44.1 | 36.8 | 35.2 |  |  |
| 2010 | 34.6 | 33.8 | 45.9 | 44.5 | 39.8 | 36.9 |  |  |
| 2011 | 33.7 | 32.9 | 44.7 | 43.3 | 38.1 | 35.5 |  |  |
| 2012 | 33.8 | 33.2 | 44.3 | 42.9 | 37.1 | 35.7 |  |  |
| 2013 | 34.4 | 34.6 | 44.8 | 42.9 | 38.0 | 36.5 |  |  |
| 2014 | 35.0 | 34.8 | 45.6 | 42.9 | 40.4 | 37.4 |  |  |
| 2015 | 34.5 | 34.8 | 45.6 | 42.7 | 40.9 | 38.3 |  |  |
|  |  |  |  |  |  |  |  |  |

Table 3.2.3.1. Summary output of the TV-survey in 3.a from 2015.

| Subarea | $\begin{aligned} & \text { AREA } \\ & \text { (Km2) } \end{aligned}$ | Number of STATIONS | Absolute MEAN DENSITY | 95\% Confidence INTERVAL | Population numbers (MILL.) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2044 | 34 | 0.319 | 0.119 | 653 |
| 2 | 1982 | 24 | 0.392 | 0.354 | 777 |
| 3 | 2462 | 36 | 0.398 | 0.143 | 980 |
| 4 | 676 | 14 | 0.630 | 0.228 | 426 |
| 5 | 670 | 12 | 0.430 | 0.380 | 288 |
| 6 | 973 | 19 | 0.541 | 0.188 | 527 |
| 7 | 1019 | 15 | 0.203 | 0.100 | 207 |
| Total | 9826 | 154 | 0.393 | 0.203 | 3857 |
|  |  |  | Harvest rate |  | 0.0205 |
| Removals 2015 (landings + dead discards**) ${ }^{\text {\% }}$ |  |  |  | $79^{*}$ |  |

* In millions
**The survival rate of discard is estimate to be $\mathbf{2 5 \%}$ (Wileman et al. 1999)


Figure 3.1.1. Nephrops Functional Units in the North Sea and Skagerrak/Kattegat region.

Illa catches, 2015.

## By landings and discards



Figure 3.2.1.1. Skagerrak (FU 3) and Kattegat (FU4): Length frequency distributions of Nephrops catches, split by catch fraction (landings and discards) and sex. Data for Denmark and Sweden combined for 2015.


Figure 3.2.2.1. Nephrops Skagerrak (FU 3): Long-term trends in landings, effort, LPUEs, and mean sizes of Nephrops.

Skagerrak (FU3) Mean sizes in Skagerrak catches


Figure 3.2.2.2. Nephrops in FU 3. Mean sizes in the catches.


Figure 3.2.2.3. Nephrops in FU 3. Danish LPUE trends.


Figure 3.2.2.4. Nephrops Kattegat (FU 4): Long-term trends in landings, effort, LPUEs.

Mean sizes in Kattegat catches


Figure 3.2.2.5. Nephrops in FU 4. Mean sizes in the catches.


Figure 3.2.2.6. Nephrops in FU 4. Danish LPUE trends.


Figure 3.2.3.2. The defined sub areas of the Nephrops stock in 3.a.



Figure 3.2.3.3. The spatial distribution of the Danish and Swedish Nephrops fishery in 2010. Left map shows VMS pings and the right map shows density of VMS pings.


Figure 3.2.3.4. Sampling locations and Nephrops burrow density in the UWTV survey in the Skagerrak and Kattegat (FU 3 and 4) in 2008 ( 26 stations, Denmark only), 2009 ( 47 stations, Denmark only), 2010 ( 72 stations, Denmark only), 2011 ( 146 stations), 2012 ( 166 stations), 2013 ( 157 stations), 2014 ( 154 stations) and 2015 ( 154 stations).


Figure 3.2.4.1 Nephrops in Area 3.a. Combined Effort for FU 3\&4


Figure 3.2.4.2 Nephrops in Area 3.a. Combined LPUE for FU 3\&4. Red dotted line shows the year at the shift in Danish management system.


Figure 3.2.4.3 Nephrops in 3.a FUs 3\&4. Catch by sex and size category in numbers and biomass.

Mean burrow density in FU3 \& 4


Figure 3.2.4.4. Mean burrow density in 3.a by year. Error bars indicate the $95 \%$ confidence intervals.

## 4 Norway lobster (Nephrops spp.) in Subarea 4 (North Sea)

### 4.1 General comments relating to all Nephrops stocks

See section 3.1

### 4.2 Nephrops in Subarea 4

Subarea 4 contains nine FUs $5,6,7,8,9,10,32,33$ and 34 . Management is applied at the scale of ICES Division through the use of a TAC and an effort regime. FU34 (The Devil's Hole) is a relatively new functional unit having been designated in 2010 (SGNepS 2010). The forecasts for FUs 6,7 and 8 were re-run in the Autumn because new information from UWTV surveys held during the summer triggered the re-opening criterion - the updated forecasts are provided in Annex 04.

## Management at ICES Subarea Level

The 2014 EC TAC for Nephrops in ICES Subarea 2.a and 4 was 15499 tonnes in EC waters (plus 1000 tonnes in Norwegian waters). For 2015, this was increased to 17843 tonnes in EC waters and 1000 tonnes in Norwegian waters.

A major change in the management of Nephrops fisheries in ICES Subarea 4 for 2016 is the introduction of the landing obligation for Nephrops fisheries in the $80-99 \mathrm{~mm}$ trawl fisheries. This does not affect the historical assessments presented here, but should affect the forecasts for 2017 fishing opportunities. A de minimis excemption for catches below the Minimum Conservation Reference Size (MCRS) of up to $6 \%$ was permitted for the fishery in Subarea 4. Where discard length frequencies and rates were available, the proportion of catch (in biomass) of animals below the MCRS was always below 6\% and therefore no change in fishing behaviour would be expected from this rule. Discards above MCRS are generally for reasons of quality (damaged/soft) and these would be expected to now be landed. Catch options therefore are presented in four categories, "wanted landings", "unwanted landings" (catch historically discarded but above MCRS), "de minimis discards" and surviving discards (as not all discards die).

The minimum landings size (MLS) for Nephrops in Subarea 4 (EC) is 25 mm carapace length. Denmark, Sweden and Norway applied a national MLS of 40 mm up to 2015 but this was changed to 32 mm from 01/01/2016.

Days-at-sea regulations and recently introduced effort allocation schemes ( $\mathrm{kW}^{*}$ day) have reduced opportunities for directed whitefish fishing. STECF 2010 stated that the overall effort (kW*days) by demersal trawls, seines and beam trawls shows a substantial reduction since 2002. However, there have also been substantial changes in the usage of the different mesh size categories by the demersal trawls. In particular there has been a sharp reduction in usage of gears with a mesh size of between 100 mm and 119 mm (targeting whitefish), but only a gradual decline in the effort of Nephrops vessels (TR2).

UK legislation (SI 2001/649, SSI 2000/227) requires at least a 90 mm square mesh panel in trawls from 80 to 119 mm , where the rear of the panel should be not more than 15 m from the cod-line. The length of the panel must be 3 m if the engine power of the vessel exceeds 112 kW , otherwise a 2 m panel may be used. Under UK legislation, when fishing for Nephrops, the cod-end, extension and any square mesh panel must be constructed of single twine, of a thickness not exceeding 4 mm for mesh sizes $70-99 \mathrm{~mm}$,
while EU legislation restricts twine thickness to a maximum of 8 mm single or 6 mm double.

Under EU legislation, a maximum of 120 meshes round the cod-end circumference is permissible for all mesh sizes less than 90 mm . For this mesh size range, an additional panel must also be inserted at the rear of the headline of the trawl. UK legislation also prohibits twin or multiple rig trawling with a diamond cod end mesh smaller that 100 mm in the North Sea south of $57 \circ 30^{\prime} \mathrm{N}$.

Official catch statistics for Subarea 4 are presented in Table 3.3.1. The preliminary officially reported landings in 2014 are around 9600 tonnes, representing a decline of over $50 \%$ from the peak observed in 2009 ( 24500 t ). The decline has been almost entirely from within the UK which typically represents $80 \%$ of the landings. Overall quota uptake for 2015 was $51 \%$.

Table 3.1.2 shows landings by FU as reported to the WG. It also shows that a small but significant proportion of the landings from Subarea 4 come from outside the defined Nephrops FUs. This value increased to nearly $10 \%$ of the total in 2009 and as a response, a new Functional Unit at the Devil's Hole (FU 34) was designated in 2011. The trends observed in the 2012 Fishers' North Sea stock survey for Nephrops are discussed in the Quality of Assessment sections for each FU but there was no update to this survey in 2016.

### 4.3 Botney Gut (FU5)

### 4.3.1 The fishery in 2014 and 2015.

Landings from FU5 have been increasing from a low point in 2009 and were the highest on record in 2015 at 1516 (more than double the 2009 landings). Over the last 15 years the national composition of the fleet fishing this FU has changed with Belgium reducing its landings and the UK increasing. Since 2004, the Netherlands and UK have consistently taken between $78-82 \%$ of the landings. Danish activity has been at a low level but erratic since 2006.

Nephrops in FU5 are caught by trawling, there is no creeling in the area.

### 4.3.2 Data Available

## Landings

Landings by country for FU 5, including Belgium, Denmark, Netherlands, Germany and UK are available since 1991 (Table 4.3.1). Landings increased from $\sim 800 \mathrm{t}$ in the early 1990s to $\sim 1200$ t in the early 2000 s, peaking at $\sim 1400$ t in 2001 . There then followed a period of general decline to a low in 2009 but landings have subsequently increased and in 2015 were 1516, the highest on record. Between 1991 and 1995, the Belgian fleet took more than $75 \%$ of the international Nephrops landings from this FU, but since then, the Belgian landings have declined drastically, and since 2006 there has been no directed Belgian Nephrops fishery by Belgian operated vessels. Some Belgian owned vessels operating as Dutch vessels have a directed fishery and trebled the landings between 2009 and 2014. Danish landings have been sporadic since 2006. In the most recent years UK and Netherlands have accounted for most of the landings from this FU, the large increase in landings 2014-2015 being driven entirely by these two fleets.

Annual discard data for 2015 were available from the Dutch self sampling program. Discard data were available for the Belgian Nephrops fleet for the period 2002-2005 but
in the absence of a directed fishery since 2006, there have been no data collection from the Belgian Nephrops landings. Three different discarding patterns were observed in the Dutch sampling. The $70-99 \mathrm{~mm}$ otter trawl fleet had a very high discard rate $(93 \%$ by number), the $70-99 \mathrm{~mm}$ beam trawl fleet reported a $2.8 \%$ discard rate by number and all other gears reported zero discarding.

## Length compositions

Length compositions in the Dutch landings are available from 2004 to 2015 with the exception of 2013 (Figure 4.3.1). Length composition for the 2015 discard data from the Dutch self-sampling program were also available. Data for 2013 were not considered of sufficient quality for inclusion due to a large SOP error (SOP=sum of products, the sum of number landed at length * weight at length should be close to the total landing biomass). Both mean sizes of males and females showed an increasing trend over time up to around 2012 but have been stable since (Table 4.3.2), although the intensity of sampling is fairly low in FU 5 and as a result samples may not be fully representative of actual removals. Between 2005 and 2009 the average numbers measured were >10 000 individuals a year, while between 2010 and 2012 the sampling measurements dropped to around 2500-3000 individuals. Sampling intensity in 2011 and 2012 was particularly low in the third quarter which is the main period of the fishery.

### 4.3.3 Natural mortality, maturity at age and other biological parameters

No analytical assessment has been performed this year.
In previous analytical assessments (see e.g. WGNEPH, 2003), natural mortality was assumed to be 0.3 for males of all ages and in all years. Natural mortality was assumed to be 0.3 for immature females, and 0.2 for mature females. Discard survival was assumed to be 0.25 for both males and females (after Gueguen \& Charuau, 1975, and Redant \& Polet, 1994).

Growth parameters are as follows:
Males: $\mathrm{L} \infty=62 \mathrm{~mm}$ CL, $\mathrm{k}=0.165$.
Immature females: $\mathrm{L} \infty=62 \mathrm{~mm} C L, \mathrm{k}=0.165$.
Mature females: $\mathrm{L} \infty=60 \mathrm{~mm} \mathrm{CL}, \mathrm{k}=0.080$, Size at $50 \%$ maturity $=27 \mathrm{~mm}$ CL.
Growth parameters have been assumed to be similar to those of Scottish Nephrops stocks with similar overall size distributions of the landings (see e.g. WGNEPH, 2003). Female size at 50\% maturity was taken from Redant (1994).

### 4.3.4 Commercial catch-effort data and research vessel surveys

Effort and LPUE data have been presented for this FU for several years as indicator indices however in 2015 it was discovered that there were serious concerns regarding the way in which effort had been reported for the Dutch, Belgian and English fleets.

- Historic Belgian effort data claimed to be from vessels targeting Nephrops however it transpired that the effort data were for the whole towed gear fleets operating in this area. It was not possible to reconstruct a new effort series in time for the 2015 assessment meeting but this should be completed in time for the 2016 assessment.
- Dutch data had always stated that effort was being reported for all vessels catching Nephrops but closer investigation showed that the majority of the
landings were made from TR2 gears whilst the effort figures were dominated by TBB gears where Nephrops are picked up as bycatch. Revised LPUE indices will be developed in time for the next assessment in 2016.
- English effort data purported to represent hours fished, however there were discovered to be a large number of inconsistent entries in the effort fields on the database. It was decided that reporting effort in hours fished was not a viable option and therefore days fishing for targeted Nephrops activity have been reported instead. A landing targeting Nephrops is defined as using 7099 mm otter trawl with at least $25 \%$ by weight of Nephrops per record. Changes to the way in which fishing activities were recorded in 2006 significantly sharply changing the level of landings and targeted reported.
- In addition to the erroneous data in the Dutch, Belgian and English data, Danish activity in the area has become sporadic with only one or two vessels prosecuting the fishery, therefore LPUE data for this sector is not used as an index of abundance.

Changes to the way in which gear is specified in the English fishery since 2014 necessitates a re-calculation of the landings and effort for the directed fishery. The basic premise of the calculation remains the same, otter trawl gears in the $70-99 \mathrm{~mm}$ category in which $>25 \%$ of the total landing comprises Nephrops, but the number of days fishing included in this categorisation has increased since the previous data extraction however there are minimal changes to the resulting LPUE. The only LPUE series considered to be an appropriate abundance proxy is the English LPUE series since 2006 (Table 4.3.3 and Figure 4.3.2.)

Effort by English vessels targeting Nephrops (targeting being classed as trips where Nephrops comprise >=25\% of landings by biomass) in FU5, has been generally falling since 2006 but was relatively stable between 2013-2015. LPUE has fluctuated without trend over the period 2006-2015.

## TV Survey in FU5 (Botney Gut / Silver Pit):

There were no new surveys in this FU since the last assessment in 2013. Details of the 2010 and 2012 surveys are given in the WGNSSK report from 2013.

## Intercatch

FU5 data were available from Intercatch for all nationalities for 2011, 2012, 2013 and 2014. Quarterly landings by metier were available for all countries fishing the functional unit. Length composition data were not available for 2013 as the sample rate was considered insufficient to raise the distributions. Discards were raised for non-sampled strata on a fleet by fleet basis, matching the Dutch raising factors for the three different gear types (70-99 Otter, 70-99 Beam and other). This approach gave an overall discard rate of $73 \%$ by number ( $57 \%$ by weight), which is substantially higher than the $25 \%$ by number assumed in previous assessments. As a bounding exercise, the non-observed fleets had discard rates set to zero resulting in a total international discard rate of $45 \%$, indicating that discard data for other fisheries is highly desirable (although with the inclusion of Nephrops into the landing obligation in 2016 this need should reduce).

### 4.3.5 Status of stock

The status of this stock is uncertain although there are no consistent signals that this stock is suffering from over-exploitation. The lack of reliable of length information on
this stock in recent years means that there is no information regarding incoming recruitment. The advent of discard data from the Dutch fleet for 2015 indicates that harvest rates have been significantly higher than previously assumed.

In previous assessments there had been concern regarding conflicting signals in LPUE series, however this was most likely due to the inclusion of non-targeted behaviours and there is now only one LPUE series presented which shows no trends.
Following the procedure outlined in section 3.1.2, an estimate of the total Nephrops grounds was used to give a likely envelope for the total abundance of Nephrops in this functional unit. Mean weight in landings component came from the 3 year average whilst the mean weight of discards and the discard proportion came from the 2015 data as they were the only ones available. Discard survival was set to zero in line with the protocol for data limited Nephrops stocks. The 2012 survey shows that density is relatively high on this ground at 0.7 burrows per metre squared.

Under the assumption that the landing obligation applies (i.e. no discarding) and that the abundance is around 0.7 burrows per metre squared, the 2016 catch advice should result in a Harvest Rate of around $3.6 \%$. Even a $20 \%$ increase in the 2016 catch advice should result in a Harvest Rate of less than 7.5\% (the lowest MSY harvest rate for the analytically assessed stocks in area 4).

| BASIS | Total Catch | Wanted CATCH | UnWanted CATCH | Dead DISCARDS | Surviving DISCARDS | 0.05 | 0.1 | 0.2 | 0.3 | 0.4 | 0.5 | 0.6 | 0.7 | 0.8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.2*2015 | 837 | 303 | 534 | - | - | 40.0\% | 20.0\% | 10.0\% | 6.7\% | 5.0\% | 4.0\% | 3.3\% | 2.9\% | 2.5\% |
| 2016 Catch advice | 1155 | 419 | 737 | - | - | 55.2\% | 27.6\% | 13.8\% | 9.2\% | 6.9\% | 5.5\% | 4.6\% | 3.9\% | 3.5\% |
| 0.3*2015 | 1256 | 455 | 801 | - | - | 60.0\% | 30.0\% | 15.0\% | 10.0\% | 7.5\% | 6.0\% | 5.0\% | 4.3\% | 3.8\% |
| $\begin{aligned} & 2016 \\ & \text { Catch adv } \\ & { }^{*} 1.2 \end{aligned}$ | 1386 | 502 | 884 | - | - | 66.3\% | 33.1\% | 16.6\% | 11.0\% | 8.3\% | 6.6\% | 5.5\% | 4.7\% | 4.1\% |
| 0.5*2015 | 2093 | 758 | 1335 | - | - | 100.1\% | 50.0\% | 25.0\% | 16.7\% | 12.5\% | 10.0\% | 8.3\% | 7.1\% | 6.3\% |
| Fmsy | 2195 | 795 | 1400 | - | - | 104.9\% | 52.5\% | 26.2\% | 17.5\% | 13.1\% | 10.5\% | 8.7\% | 7.5\% | 6.6\% |
| $\begin{aligned} & 0.8^{*} 10 \mathrm{yr} \\ & \text { av } \end{aligned}$ | 2304 | 834 | 1469 | - | - | 110.1\% | 55.1\% | 27.5\% | 18.4\% | 13.8\% | 11.0\% | 9.2\% | 7.9\% | 6.9\% |
| $0.8^{*} 2016$ land advice | 2472 | 895 | 1576 | - | - | 118.2\% | 59.1\% | 29.5\% | 19.7\% | 14.8\% | 11.8\% | 9.8\% | 8.4\% | 7.4\% |
| 2016 landings advice | 2880 | 1043 | 1837 | - | - | 137.7\% | 68.8\% | $34.4 \%$ | 22.9\% | 17.2\% | 13.8\% | 11.5\% | 9.8\% | 8.6\% |
| 10 year av | 3089 | 1119 | 1970 | - | - | 147.7\% | 73.9\% | 36.9\% | 24.6\% | 18.5\% | 14.8\% | 12.3\% | 10.6\% | 9.2\% |
| $0.8 * 2015$ | 3348 | 1213 | 2136 | - | - | 160.1\% | 80.0\% | 40.0\% | 26.7\% | 20.0\% | 16.0\% | 13.3\% | 11.4\% | 10.0\% |
| 2015 | 4186 | 1516 | 2670 | - | - | \#\#\#\#\# | \#\#\#\#\# | 50.0\% | 33.4\% | 25.0\% | 20.0\% | 16.7\% | 14.3\% | 12.5\% |

Under the scenario in which the landing obligation applies but discarding is permitted on a di minimis basis (i.e. catch below MCRS is permitted to be discarded up to $6 \%$ of total catch weight), the same $20 \%$ increase in total catch advice from the 2016 value results in a harvest rate below the $7.5 \%$ level.

| BASIS | Total CATCH | Wanted Catch | Unwanted <br> Catch <br> $>$ MCRS | De minimis DISCARDS <MCRS | Surviving DISCARDS | 0.05 | 0.1 | 0.2 | 0.3 | 0.4 | 0.5 | 0.6 | 0.7 | 0.8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.2*2015 | 837 | 303 | 500 | 34 | 0 | 40.0\% | 20.0\% | 10.0\% | 6.7\% | 5.0\% | 4.0\% | 3.3\% | 2.9\% | 2.5\% |
| 2016 Catch advice | 1156 | 419 | 691 | 46 | 0 | 55.2\% | 27.6\% | 13.8\% | 9.2\% | 6.9\% | 5.5\% | 4.6\% | 3.9\% | 3.5\% |
| 0.3*2015 | 1256 | 455 | 751 | 50 | 0 | 60.0\% | 30.0\% | 15.0\% | 10.0\% | 7.5\% | 6.0\% | 5.0\% | 4.3\% | 3.8\% |
| $\begin{aligned} & 2016 \\ & \text { Catch adv } \\ & { }^{*} 1.2 \end{aligned}$ | 1386 | 502 | 828 | 56 | 0 | 66.3\% | 33.1\% | 16.6\% | 11.0\% | 8.3\% | 6.6\% | 5.5\% | 4.7\% | 4.1\% |
| 0.5*2015 | 2093 | 758 | 1251 | 84 | 0 | 100.1\% | 50.0\% | 25.0\% | 16.7\% | 12.5\% | 10.0\% | 8.3\% | 7.1\% | 6.3\% |
| Fmsy | 2195 | 795 | 1312 | 88 | 0 | 104.9\% | 52.5\% | 26.2\% | 17.5\% | 13.1\% | 10.5\% | 8.7\% | 7.5\% | 6.6\% |
| $\begin{aligned} & 0.8^{*} 10 \mathrm{yr} \\ & \text { av } \end{aligned}$ | 2304 | 834 | 1377 | 93 | 0 | 110.1\% | 55.1\% | 27.5\% | 18.4\% | 13.8\% | 11.0\% | 9.2\% | 7.9\% | 6.9\% |
| $\begin{aligned} & 0.8^{*} 2016 \\ & \text { land } \\ & \text { advice } \end{aligned}$ | 2472 | 895 | 1477 | 99 | 0 | 118.2\% | 59.1\% | 29.5\% | 19.7\% | 14.8\% | 11.8\% | 9.8\% | 8.4\% | 7.4\% |
| 2016 landings advice | 2880 | 1043 | 1721 | 116 | 0 | 137.7\% | 68.8\% | 34.4\% | 22.9\% | 17.2\% | 13.8\% | 11.5\% | 9.8\% | 8.6\% |
| 10 year av | 3090 | 1119 | 1847 | 124 | 0 | 147.7\% | 73.9\% | 36.9\% | 24.6\% | 18.5\% | 14.8\% | 12.3\% | 10.6\% | 9.2\% |
| 0.8*2015 | 3349 | 1213 | 2002 | 135 | 0 | 160.1\% | 80.0\% | 40.0\% | 26.7\% | 20.0\% | 16.0\% | 13.3\% | 11.4\% | 10.0\% |
| 2015 | 4186 | 1516 | 2502 | 168 | 0 | 200.1\% | 100.1\% | 50.0\% | 33.4\% | 25.0\% | 20.0\% | 16.7\% | 14.3\% | 12.5\% |

### 4.3.6 Management considerations for FU 5.

The North Sea TAC is not thought to be restrictive for the fleets exploiting this stock. Given the paucity of metrics available for monitoring stock development, the exploitation of this stock should monitored closely.

### 4.4 Farn Deeps (FU6)

An updated assessment based on the 2015 TV survey is available in Annex 04.

### 4.4.1 Fishery in 2014 \& 2015

Since the beginning of the time-series, the UK fleet has accounted for virtually all landings from the Farn Deeps (Table 4.4.1). The Farn Deeps fishery is essentially a winter fishery commencing in September and running through to March, hence the 2015 data comprise the end of the 2014-2015 fishery and the start of the 2015-2016 fishery.

Landings in 2014 were 2503t, close to the 10 year average whilst landings in 2015 dropped to 1371, the lowest since 2008 and the second lowest on record (Figure 4.4.1). The majority of this reduction occurred in the second half of 2015, the winter fishery of 2015-2016 being particularly poor.

The introduction of the buyers and sellers legislation in 2006 means direct comparison with previous years should be viewed with caution because the suspected resulting improvement in reporting levels will have created a discontinuity in the data.

Directed effort (i.e. days fishing by vessels fishing with Nephrops gears) from English vessels during 2015 declined from the 2014 level, particularly for the $<10 \mathrm{~m}$ sector, although the $10-15 \mathrm{~m}$ sectors also declined substantially. Only in the $>15 \mathrm{~m}$ sector did effort remain relatively constant.

Historically the fishery has been prosecuted by a combination of local English boats (smaller vessels undertaking day-trips) and larger vessels from Scotland with occasional influxes of effort by Northern Irish vessels. The total number of vessels in the fishery has generally been decreasing since 2007 although 2013 and 2014 saw increases again in all sectors (Figure 4.4.2). The majority of the dynamic in fleet size is due to Scottish boats, likely to be a response of vessels moving away from reduced catch rates in FU7.

ICES Advice in 2015
The last assessment of Nephrops in FU6 was in 2014
ICES advises that when the MSY approach is applied, catches in 2016 (assuming a landing obligation applies) should be no more than 738 tonnes. If this stock is not under the EU landing obligation in 2016 and discard rates do not change from the average (2012-2014), this implies landings of no more than 680 tonnes.

In order to ensure the stock in this FU is exploited sustainably, management should be implemented at the functional unit level.

Management is at the ICES Subarea level as described at the beginning of Section 3.3.

### 4.4.2 Assessment

## Review of the 2015 assessment

The assessment has been performed correctly with no deviations from the standard procedure for this stock. The update assessment gives a valid basis for advice.

## Data available

## Catch, effort and research vessel data

Three types of sampling occur on this stock, landings sampling, catch sampling and discard sampling providing information on size distribution and sex ratio. The sampling intensity is considered to be generally good although concerns regarding the sampling levels of tail (as opposed to whole) landings has resulted in the catch and landings distributions being estimated from the monthly catch samples, supplemented by the discard sampling. The use of landings sampling where the tailed portion of the catch is under-represented would upwardly bias the estimate of landing lengths.

## Discards

The procedure used to estimate discards changed in 2002. The methods are described in detail in the Stock Annex. Discarding practice varies considerably between vessels in any given period but there is no significant trend in the computed discard ogives (Figure 4.4.4) hence the use of a fixed discard ogive on the catch length distributions since 2002. The Benchmark meeting in 2013 concluded that the historical assumption of $0 \%$ discard survival was no longer applicable as a significant proportion of catch sorting now takes place at sea. For day-boats, the first haul of the day will generally be sorted on the fishing grounds whilst the second haul will be sorted whilst steaming back to port (and therefore passing over habitat unsuitable for Nephrops. Discarding practice for multi-day boats will generally result in discards returning to suitable sediment. The conclusion was therefore that although the full $25 \%$ survival assumed in other FUs was not likely to be applicable a $15 \%$ survival rate was a reasonable estimate for this FU.

## Length Frequency

There is a clear change in length frequencies around 2007 with much lower contributions from the smaller (discarded) size classes (Figure 4.4.7). This may reflect an improvement in selectivity by the fleet or alternatively a decrease in recruitment levels. There is a decrease in the overall level of TV survey around the same time indicating that this change in length distribution may at least partly reflect a reduction in the level of recruitment.

A bi-modal length frequency distribution for landed females had been present since 2009 and become steadily more pronounced until 2014. This could be the result of a large year class, but a similar phenomenon is not observed in the male part of the population, in fact the mean size in the males decreased in 2012 and 2013. The mean annual increment of the female second mode is only around 2 mm whereas interannual growth would be expected to be more and therefore year class strength is unlikely to be the cause of this feature. The predominance of large females in the catches means they were foraging for food on the surface at a time when they would have been expected to be brooding eggs within their burrows. Given that there are very few males of similar size appearing in the catches it is possible that there is a physical size differential constraint in mating patterns of Nephrops. This may either be an inability of the males to successfully transfer spermatophores, or alternatively large females may be able to resist the (usually quite aggressive) approaches of the smaller males when they try to mate with large females. The reduction in the bi-modal nature of the female length distribution in 2015 implies a lower relative availability of females at larger sizes and may indicate a better spawning success. The alternative hypothesis is that this part of the population was removed by the fishery and would therefore suggest continued low recruitment.

There is therefore considerable concern that this stock is likely to be suffering from reduced recruitment and may continue to do so for at least the next two years (assuming that recruits enter the fishery between age 2 and 3 ). Whether the change in proportion of large mature females in the landings is a result of improved mating conditions remains to be seen and will not be evident for a further 12-24 months in the form of improved recruitment.

## Effort and LPUE

The metric of fishing effort (hours fished by Nephrops targeting vessels) produced for 2014 using the standard raising process was sharply different to the 2013 metric. On closer inspection it transpired that changes to reporting mechanisms (more uptake of e-log systems) had caused a discontinuity in how data were reported. Further analysis of historic data also highlighted serious inconsistency in the way that fishing hours had been reported back through time. The number of days fished is considered to be more reliably captured by the official statistics and therefore the effort metrics were reworked from 2000 onwards.

The way in which data regarding both landings and effort were collected within the UK changed in 2006 (Buyers and Sellers legislation) which had a noticeable change in the level of reported metrics. Comparison between these two time periods is therefore inadvisable.

Directed effort fell for the under 10m sector has remained constant from 2006 onwards, whilst for the larger vessels this dropped from 2006. 2006 saw a large influx of larger vessels from other areas of the UK including Scotland and Northern Ireland. There has
been an increase in the number of Scottish boats, particularly the large $>15 \mathrm{~m}$ sector in 2013 and 2014. (Figure 4.4.2).

The use of LPUE as an index of stock abundance for Nephrops is confounded by changes in availability of Nephrops to fishing gears depending upon environmental factors such as tide and light levels, plus changes to emergence behaviour induced by mating and predator avoidance. There is a general level of agreement between LPUE for the different gears and a decline in stock abundance from around 2006 to 2008.

Effort and LPUE show distinct differences between vessel size classes, twin-rig being more efficient for a size class of vessel than single-rigs (figure 4.4.5). There remain some consistency issues between periods pre and post 2006, but despite this there appears to be a difference in the trajectory of LPUE between the vessel size classes. Small ( $<10 \mathrm{~m}$ ) single-rig vessels have seen a sharp drop from $\sim 300 \mathrm{~kg}$ per day in 2006 to around 150 kg per day in 2014 and this level had remained fairly constant between 2009 and 2014, vessels larger than this on the other hand appeared to be experiencing increasing LPUE, particularly for the $>15 \mathrm{~m}$ vessel sector. This may represent a spatial difference in stock development as the smaller vessels are restricted to more inshore areas, however there may be some fleet changes (larger vessels) or reporting changes (issues with data from e-logs) which are driving some of these differences. For 2015 however, all fleet sectors experienced their lowest recorded catch rates.

Traditionally, males tend to predominate the landings, averaging about 70\% (range $64 \%-79 \%$ ) by biomass in the period 1992-2005. Towards the end of the fishing season (February-March) there is usually an increase in female availability as mature females emerge from their burrows having released their eggs. There has been a marked change in the seasonal pattern of sex-ratio for Farn Deeps Nephrops since the winter of 2005. Prior to this the ratios were generally smooth with small ( $\sim 10 \%$ ) seasonal fluctuations, but since then the fishery has observed very large swings, with whole years being dominated by landings of females (2006, 2010, 2012 and 2013-2014, Figure 4.4.4). The sex ratio for 2015 returned to a generally male dominated fishery and can be explained by the lack of large females in the catches (figure 4.4.7).

Effort in the 2014-2015 winter fishery was markedly lower than the same period 12 months previously but no lower than that observed in the early 2000s when abundance was estimated to be much higher. The relative strength of effort within a season (i.e. the fourth quarter compared to the first quarter) fluctuates without trend.

Female LPUE in the fourth quarters of 2000, 2006, 2009, 2001 and 2013 have been higher than one might expect given that they are supposed to have reduced availability due to egg-brooding.

## UWTV

Underwater TV surveys of the Farn Deeps grounds have been conducted at least once in each year from 1996 onwards. Initially there were two surveys, one in the autumn preceding the fishery and one in the spring immediately after the fishery, however only the autumn survey has continued. In 2013 the UWTV survey of the Farn Deeps was carried out in June for operational reasons. The potential change in survey timing was presented to the Benchmark meeting in 2013 and a report of the discussions is contained within the report of that group (ICES 2013). The following points were considered.

- There is practically no targeted Nephrops fishing between May and September; therefore there is minimal scope for fishery induced changes in stock abundance between the new survey time and the previous October timing.
- There are no migrations of animals to consider, Nephrops are on (or rather in) the ground all year round. What affects their availability to the fishery is their emergence behaviour, which does not affect the ability to count the burrows.
- The only factor which may affect the burrow density between June and September is any seasonality in the creation of new burrow complexes by juveniles. There are no data regarding the timing of burrow creation with which to make an informed judgement as to whether this is likely to be an important effect.

The Benchmark group concluded therefore that the FU 6 TV series should continue to be regarded as a single series and that where possible work should be undertaken to investigate if the survey timing change was likely to have affected results.

A time series of indices is given in Figure 4.4.9 and table 4.4.5. The procedure used to work up the TV survey has been changed in 2011. The original survey design was a random-stratified design where the ground was split into regular boxes with stations randomly placed within. At a later stage additional stations were inserted into areas of high density to better define them, however this was not accounted for in the process of estimating overall abundance and therefore the higher density of stations in highdensity Nephrops areas will have biased the estimate upwards. In addition, the distance covered by the TV sledge was determined by assuming a straight-line between the start and finish positions of the vessel. Since 2007, GPS logging of the position of the vessel and the sledge (via a Hi-Pap beacon) at short intervals ( $\sim 5$ seconds) has enabled a considerably more robust estimate of viewed distance to be made. The abundance estimate is now made using a geostatistical procedure in which the spatial position of the burrow density estimates are first fitted by a semi-variogram model and then a 3D surface of burrow density is created using Kriging on a $500 \mathrm{~m} * 500 \mathrm{~m}$ grid. Uncertainty estimation of the overall abundance estimate is performed by bootstrapping the counts, refitting the semi-variogram and re-estimating the surface. Uncertainty estimates are typically $2 \%$, much lower than the previous estimates which ignored spatial structure to a large degree. Figure 4.4 .10 shows the final maps along with the abundance estimates. The TV survey in 2009 was hampered by a period of poor weather and low visibility which coincided with the surveying of the areas traditionally associated with the highest densities (fishing vessels were working this area at the time of survey and consequently disturbing the sediment). The spatial pattern of burrow density is similar through time with the highest density ground running along the eastern edge of the mud-patch.

## Intercatch

All data for 2015 were entered onto Intercatch. Landings data by fleet were provided by Scotland, England, Denmark, Belgium and the Netherlands, whilst England provided length distributions for landings and discards by fleet where available.

Discard ratios for all unsampled fleets were raised on the combined annual data from England. Quarterly length distributions were imported for England which represetned $84 \%$ of the landings. Consequently, length frequencies for the remaining metiers were generated from the pooled data (i.e. irrespective of metier or quarter) for both landing and discard components.

## Natural mortality, maturity at age and other biological parameters

Biological parameter values are included in the Stock Annex which was updated at the 2013 benchmark.

## Exploratory analyses of RV data

A comprehensive review of the use of underwater TV surveys for Nephrops stock assessment was undertaken by WKNeph (ICES 2009). This covered the range of potential biases resulting from factors including edge effects, species mis-identification, burrow occupancy. Cumulative bias-correction factors were estimated for each FU and for FU6 the bias correction factor is 1.2 meaning that the raw counts from the TV survey are likely to overestimate densities of Nephrops by $20 \%$. The correction factor is therefore applied to the raw counts to arrive at the absolute abundance index. Estimates of absolute burrow density total abundance estimates (with confidence estimates) are given in Table 3.3.2.4.

## Final Assessment

The estimated abundance in 2015 was 549 million individuals ( $95 \%$ confidence interval of $\pm 13$ million), significantly below the 2007 estimate used as MSY $B_{\text {trigger }}$ ( 858 million). The estimated harvest rate for 2015 was $11 \%$, well above the MSY proxy level of $8.1 \%$ but lower than the 2014 Harvest Rate of 13\%.

The stock therefore remains in a vulnerable state. The dominance of large females in the landings again for the 2012-2014 fishery suggests that they had not successfully mated and therefore there remains the potential for poor recruitment for 2016 and 2017 (recruits to the fishery are estimated to be $\sim 2-3$ years old)

### 4.4.3 Historical stock trends.

The time series of TV surveys is 14 consecutive years although the new geostatistical method has only been applied retrospectively to 2007. Whilst there is expected to have been a small over-estimation of abundance using the previous technique it is likely that the reduction in stock abundance observed between the two periods of estimation procedure is real.

Estimates of historical harvest ratio (the proportion of the stock which is removed) range from $6.1 \%$ to $25.5 \%$ (Table 4.4.5). The harvest ratio jumped from around $12 \%$ in $2004-2005$ to $25.5 \%$ in 2006 when the new reporting legislation came in. The harvest rate has only been below the MSY level once in 13 years.

### 4.4.4 MSY considerations

Considerations for setting Harvest Ratios associated with proxies for $\mathrm{F}_{\text {msy }}$ for Nephrops are described in ICES, WGNSSK, 2010, section 1.

- Average density in the stock is at a medium level, above the level of the FU 7 but below that of FU 8 .
- Density has varied through time but does not appear to undergo large scale interannual fluctuations. Spatially there is a good degree of consistency in the pattern of high and low density between the years.
- Estimated growth rates are at a moderate level although the data supporting them are quite old. Natural mortality estimates are standard.
- The fishery in the Farn Deeps is a winter fishery (October - March) with typically male dominated catches. The intra-annual pattern of sex ratios in
the catches has changed in 2006 and 2009 possibly due to sperm limitation leading to more mature but unfertilised females being available to the fishery. This may lead to reduced recruitment to the fishery.
- Although the time series of observed harvest rates is relatively short, there has been a fair degree of fluctuation (7-25\%). The observed harvest rate is, of course, confounded by the change in reporting levels considered to have occurred around 2006. The average harvest rate since 2006 is $15.7 \%$ which is above the most recent estimate of $\mathrm{F}_{\text {max }}$ for males.

The following table shows the mean F , implied harvest rate and resulting spawner per recruit values (expressed as a percentage of virgin) for the range of $\mathrm{F}_{\mathrm{msy}}$ proxies suggested for Nephrops stocks. These values were last recalculated in 2013 using a length cohort analysis model (SCA, see ICES, WKNep 2009) on the combined length frequencies for 2010-2012. The model fit to the data (Figure 3.3.2.11) is reasonable but the increasing bi-modality of the length frequency observed in the females over the past 4 years does violate model assumptions and the model under-predicts the landings of larger females.

|  |  | Fbar 20-40 mm |  | HARVESt Rate | \% Virgin SPAWNER PER RECRUIT |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  | Female | Male |  | Female | Male |
| F0.1 | Comb | 0.09 | 0.09 | $8.7 \%$ | $47.52 \%$ | $32.11 \%$ |
| F0.1 | Female | 0.16 | 0.16 | $14.0 \%$ | $32.63 \%$ | $18.26 \%$ |
| F0.1 | Male | 0.07 | 0.07 | $7.1 \%$ | $53.02 \%$ | $38.50 \%$ |
| F35\% | Comb | 0.12 | 0.12 | $11.1 \%$ | $39.98 \%$ | $24.50 \%$ |
| F35\% | Female | 0.17 | 0.17 | $15.2 \%$ | $34.82 \%$ | $16.64 \%$ |
| F35\% | Male | 0.16 | 0.16 | $8.1 \%$ | $57.17 \%$ | $34.88 \%$ |
| Fmax | Comb | 0.17 | 0.17 | $15.3 \%$ | $34.58 \%$ | $16.48 \%$ |
| Fmax | Female | 0.29 | 0.29 | $21.6 \%$ | $22.22 \%$ | $9.47 \%$ |
| Fmax | Male | 0.12 | 0.12 | $11.6 \%$ | $44.70 \%$ | $23.73 \%$ |

The default Harvest Rate suggested for Nephrops is the combined sex F35\%SpR. The effects of sperm limitation appear to have been a factor in the recent development of this stock. There are signs that this stock may be in a period of lower productivity and so a harvest rate which gives greater protection to the spawning potential of males would be advisable. The group therefore recommends moving the $\mathrm{F}_{\text {msy }}$ proxy to the harvest rate equivalent to $\mathrm{F} 35 \%$ on males for this stock (8.1\%).

WGNSSK suggests the absolute abundance index from the TV survey as observed in 2007 (i.e. the first year when the stock was considered to be depleted in the recent series) should become a proxy for $\mathrm{B}_{\text {trigger }}\left(\mathrm{B}_{\text {trigger }}=858\right.$ million).

## Short term forecasts.

Catch and landing predictions for 2017 are given in the text table below. This assumes that the absolute abundance estimate made in June 2015 is relevant to the stock status for 2017. The ICES MSY approach dictates that where the stock status is below the trigger point, the maximum advised fishing rate should be the MSY rate adjusted by the ratio of the current stock status to the Btrigger level. For 2017 this gives

HR2017 $^{2}=$ HRMsy (8.12\%) * Abundnace 2015 (569) / Btrigger (858) = 5.3\%

Recently, to account for the landings obligation coming into force for Nephrops in 2016, the projected amount of discards (now referred to as unwanted catches) have been added to the catch options table. The advice given for 2017 considers three different scenarios: 1. Landings obligation applying for Nephrops with no discarding allowed; 2. Nephrops discarding is allowed to continue as before 2016; 3. Landings obligation with de minimis exemption applying for Nephrops with 6\% discarding (by weight) under the minimum conservation reference size (MCRS, 25 mm in the North Sea) allowed. Three catch options tables are provided to account for each of these scenarios.

Under scenario 1, all catch is assumed to be landed, no discards will survive and therefore the harvest rate is assumed to include all catch and not only landings plus dead discards. Unwanted catches (by number) are calculated using data from the on-board observer sampling programme. This value is multiplied by the mean weight in discards to obtain the projected discard weight. A column is also included to show expected landings (referred to as wanted catches) under a landings obligation (Total Catches $=$ Wanted Catches + Unwanted Catches). The total catches calculated in this way are lower than those calculated previously based on Landings + Surviving discards + Dead discards.

Under scenario 2, the catch options table includes surviving discards, in the same format as of the previous advice. This is to account for the possibility that Nephrops qualifies for a high survival exemption in which case a landings obligation would not be applicable. Discards survival for Nephrops in FU6 is assumed to be $15 \%$.

A de minimis exemption of $6 \%$ discards by weight below MCRS (Scenario 3) has been applied in the North Sea since the $1^{\text {st }}$ January 2016 and therefore, a catch options table accounting for a continuation of this rule in 2017 has been considered for the first time in the 2016 WG. The main difference from scenario 2 is that, under a de minimis exemption, if discard patterns remain unchanged, some unwanted animals above MCRS (these are typically soft animals with no commercial value) will have to be landed. As such, the catch options under this scenario include a new column for unwanted catch above MCRS (animals that would have been previously discarded) as this is not expected to be taken as landings. As all discarded animals are below MCRS under this assumption, the predicted weight of discards (dead + surviving) is lower than in scenario 2 ( $15 \%$ survival rate is still assumed).

| VARIABLE | VALUE | SOURCE | Notes |
| :--- | :--- | :--- | :--- |
| Abundance in TV <br> assessment | 565 | ICES (2016a) | UWTV 2015 |
| Mean weight in <br> landings | 28.96 | ICES (2016a) | Average 2013-2015 |
| Mean weight in <br> discards | 11.098 | ICES (2016a) | Average 2013-2015 |
| Mean weight in <br> unwanted catch >MCRS | 13.63 | ICES (2016a) | Average 2013-2015 |
| Mean weight in <br> unwanted catch <MCRS | 6.765 | ICES (2016a) | Average 2013-2015 |
| Discard rate (total) | $24.56 \%$ | ICES (2016a) | Average 2013-2015 (proportion by <br> number) |
| Discard rate (>MCRS) | $14.14 \%$ | ICES (2016a) | Average 2013-2015 (proportion by <br> number) |
| Discard rate (<MCRS) | $10.45 \%$ | ICES (2016a) | Average 2013-2015 (proportion by <br> number) |
| Discard survival rate | $25 \%$ | ICES (2016a) | Only applies in scenarios where <br> discarding is allowed. |
| Dead discard rate (total) | $21.67 \%$ | ICES (2016a) | Average 2013-2015 (proportion by <br> number), only applies in scenarios <br> where discarding is allowed. |
| Dead discard rate <br> (<MCRS) | $9.03 \%$ | ICES (2016a) | Average (proportion by number) <br> 2013-2015, only applies in scenarios <br> where when discarding allowed for <br> de minimus exemptions. |

Nephrops in FU6: Catch options assuming the landing obligation applies.

|  | Total CATCH | WANTED <br> CATCH* $^{*}$ | UNWANTED <br> CATCH* $^{*}$ | HARVEST <br> RATE** |
| :--- | :--- | :--- | :--- | :--- |
| FMSY ApproachComb | 791 | 769 | 703 | 66 |
| FmsyLower | 1036 | 1007 | 921 | 86 |
| F0.1Male | 1052 | 1023 | 935 | 87 |
| FmsyUpper | 1183 | 1151 | 1052 | 98 |
| F35\%Male $=$ FMSY | 1201 | 1168 | 1068 | 100 |
| F0.1Comb | 1284 | 1248 | 1142 | 107 |
| F35\%Comb | 1648 | 1602 | 1465 | 137 |
| FmaxMale | 1716 | 1668 | 1526 | 143 |
| F0.1Female | 2074 | 2017 | 1844 | 172 |
| F35\%Female | 2247 | 2185 | 1998 | 187 |
| FmaxComb | 2263 | 2201 | 2012 | 188 |
| Fcurrent | 2530 | 2460 | 2249 | 210 |
| FmaxFemale | 3195 | 3107 | 2841 | 266 |

Nephrops in FU6: Catch options assuming that discarding continues at historic patterns.

|  | Total <br> CATCH | DeAD <br> REMOVALS | LANDINGS | DEAD <br> DISCARDS | SURVIVING <br> DISCARDS | HARVEST <br> RATE* |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | L+DD+SD | L+DD | L | DD | SD | for <br> L+DD |
| FMSY Approach | 759 | 754 | 678 | 60 | 16 | 5 |
| FmsyLower | 994 | 987 | 887 | 78 | 22 | 7 |
| F0.1Male | 1010 | 1002 | 901 | 79 | 22 | 7 |
| FmsyUpper | 1136 | 1128 | 1014 | 89 | 25 | 8 |
| F35\%Male $=$ FMSY | 1153 | 1145 | 1029 | 91 | 25 | 8 |
| F0.1Comb | 1232 | 1224 | 1100 | 97 | 27 | 9 |
| F35\%Comb | 1582 | 1570 | 1412 | 125 | 34 | 11 |
| FmaxMale | 1647 | 1635 | 1470 | 130 | 36 | 12 |
| F0.1Female | 1991 | 1976 | 1777 | 157 | 43 | 14 |
| F35\%Female | 2157 | 2141 | 1925 | 170 | 47 | 16 |
| FmaxComb | 2172 | 2157 | 1939 | 171 | 47 | 16 |
| Fcurrent | 2428 | 2411 | 2167 | 191 | 53 | 18 |
| FmaxFemale | 3067 | 3045 | 2737 | 241 | 66 | 22 |

Nephrops in FU6: Catch options assuming the landing obligation applies with the di minimis rules for animals below MCRS.

| BASIS | Total CATCH | Dead REMOVALS | LANDINGS | UnWANTED >MCRS* | Dead <br> DISCARDS <br> <MCRS | Surviving DISCARDS | Harvest RATE** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | L+DD+SD | L+DD | L | L | DD | SD | for L+DD |
| FMSY Approach | 759 | 754 | 678 | 60 | 16 | 5 | 5.35\% |
| FmsyLower | 994 | 987 | 887 | 78 | 22 | 7 | 7.00\% |
| F0.1Male | 1010 | 1002 | 901 | 79 | 22 | 7 | 7.11\% |
| FmsyUpper | 1136 | 1128 | 1014 | 89 | 25 | 8 | 8.00\% |
| F35\%Male = FMSY | 1153 | 1145 | 1029 | 91 | 25 | 8 | 8.12\% |
| F0.1Comb | 1232 | 1224 | 1100 | 97 | 27 | 9 | 8.68\% |
| F35\%Comb | 1582 | 1570 | 1412 | 125 | 34 | 11 | 11.14\% |
| FmaxMale | 1647 | 1635 | 1470 | 130 | 36 | 12 | 11.60\% |
| F0.1Female | 1991 | 1976 | 1777 | 157 | 43 | 14 | 14.02\% |
| F35\%Female | 2157 | 2141 | 1925 | 170 | 47 | 16 | 15.19\% |
| FmaxComb | 2172 | 2157 | 1939 | 171 | 47 | 16 | 15.30\% |
| Fcurrent | 2428 | 2411 | 2167 | 191 | 53 | 18 | 17.10\% |
| FmaxFemale | 3067 | 3045 | 2737 | 241 | 66 | 22 | 21.60\% |

### 4.4.5 BRPs

Suggestions for proxies of biological reference points are shown in the catch option table and discussed in 4.3.4.

### 4.4.6 Quality of assessment

Changes to the legislation regarding the reporting of catches in 2006 means that the levels of reported landings from this point forward are considered to better reflect the true landings and hence effort input into this fishery. This does mean that comparison of LPUE with previous years is inadvisable and the independence of the final assessment from these data is likely to continue for some time.

The length and sex compositions arising from the land-based catch sampling programme are considered to be representative of the fishery. Estimates of discarded and retained length frequencies arising from the discard sampling programme are also considered robust since 2002.

The TV survey in this area has a high density of survey stations compared to other TV surveys and the abundance estimates are generally considered robust. There is greater uncertainty in the index for 2009 due to the absence of stations in the higher density areas which may result in an over-estimate of the magnitude of the decline for this year.

The spatial distribution of the 2014 survey results continues the pattern observed in other years, although at a lower overall level. The spine of high density on the western edge of the ground remains a regular feature. The main features of the survey series are peaks in abundance 2001 and 2005, with reasonably constant series since 2007.

The only harvest rate observed to be below the MSY level was in 2008, the mean harvest rate of $15.2 \%$ is almost double the MSY level.

Without suitable controls on the movement of effort between Functional Units there is nothing to prevent the effort in 2016 continuing to inflict fishing mortality above the F35\%SprR level and indeed above the level of $\mathrm{F}_{\text {max. }}$. Prior to the introduction of "Buyers and Sellers" legislation in 2006 reporting rates are considered to have been low and hence the estimated Harvest Ratios prior to 2006 are also likely to have been underestimated.

### 4.4.7 Status of stock

The 2015 TV survey indicates the stock continues to be in a depleted state and is in further decline, below the level of MSY Btrigger with harvest rates well in excess of the MSY advised rate. There are no indications of strong recruitments coming into the stock and the appearance of large females appearing in the catches up to 2014 suggests that the situation may continue for some years yet.

### 4.4.8 Management considerations

The WG, ACFM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level and management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort were compatible and in line with the scale of the resource.

Decreases in abundance in other FUs (i.e. Firth of Forth and the Fladen grounds) may raise the risk of higher effort being deployed in this FU. The high cost of fuel combined with the relative coastal proximity of this ground makes fishing this Functional Unit a
relatively attractive proposition and additional fishing effort would be inadvisable given the current low level of the stock.

During the December council meeting in 2015, the UK undertook to instigate a package of national technical measures in an attempt to bring the level of fishing mortality down. The measures to be put in place in time for the winter fishery in 2016 include:

- Vessel owners will be required to use a minimum mesh size of 90 mm using single twine of 5 mm .
- The use of a lifting bag will continue to be permitted
- Only single-rig vessels of 350 kW ( 476 hp ) or less will be permitted to fish within 12 nm of the coast.
- Multi-rig vessels (vessels with three or more rigs) will be prohibited from operating within the Farn Deeps. Twin rig vessels will be permitted to operate outside 12 nm .
- No vessel will be permitted to use gear with more than one codend per rig.
- The Farn Deeps will be defined as ICES rectangle 38E8, 38E9, 39E8, 38E9, 40 E 8 and 40E9.

The majority of the impact of these measures is expected to be a deterrent effect in terms of the number of vessels using the ground. The next ICES assessment in 2017 will have a partial year's data where these rules are in place and therefore will be in a limited position to evaluate their impacts. These rules also only apply to UK registered vessels, however the UK removals from FU6 are in excess of $95 \%$ of the historic landings.

### 4.5 Fladen Ground (FU7)

### 4.5.1 Ecosystem aspects

The Fladen Ground (Functional Unit 7) is located towards the centre of the Northern North Sea off the east coast of Scotland (Figure 3.1.1.). This region is characterised by an extensive area of mud and muddy sand, and hydrographic conditions include a large scale seasonal gyre which develops in the late spring over a dome of colder water.

Owing to its burrowing behaviour, the distribution of Nephrops is restricted to areas of mud, sandy mud and muddy sand. Within the Fladen Ground FU these substrates are distributed more or less continuously over a very large area (approx. $30000 \mathrm{~km}^{2}$ ). Figure 4.5 .5 shows the distribution of sediment in the area. Sandy mud and muddy sand are the dominant sediment types, with patches of mud in the south west area of the FU. Numerous fish species occur in in the same area as Nephrops with demersal fish more prevalent in the northern area. In the softest areas of mud, Pandalus borealis is also found.

### 4.5.2 The Fishery in 2015

The Nephrops fishery at Fladen is the largest in the North Sea and is mainly prosecuted by UK (Scotland) vessels (1774 tonnes in 2015), with England taking 4 tonnes and Denmark 8 tonnes (Table 4.5.1). Around 80 vessels participated in the Fladen fishery at various times throughout the year. The majority are Scottish vessels fishing out of and landing to Fraserburgh and Peterhead. Catch consisted of Nephrops, haddock, whiting, cod, monkfish and megrim. A number of vessels have installed freezer capabilities to enable longer trips, but the average trip is around seven days. The fishery is seasonal and the fleet nomadic, moving between Fladen, Moray Firth, Firth of Forth, Devil's

Hole, Farn Deeps and west coast of Scotland according with the time of the year and catch rates. Fishing in 2015 was generally poor in Fladen with landings reaching its lowest figure since the late eighties. A reduced number of trips were registered in 2015 with large areas of FU 7 not visited, mostly to the north of the ground. In the second quarter of 2015, low catch rates lead to a significant exodus from the Fladen grounds to the Firth of Forth, Farn Deeps and the west coast of Scotland. Information on the fishery suggests that due to poor fishing in the Minches, some vessels moved further through the west to the South of England, fishing off the Scilly Islands (FU 20-21) between April and July. Some vessels also spent some time during summer in the Silver Pits (FU 5). The fishery in Fladen improved slightly in the second half of 2015 when most landings took place, but remained low compared with recent years.

Most vessels fishing in FU 7 traditionally use twin rigs with $80 / 90 \mathrm{~mm}$ mesh. Recently, to reduce catches of whitefish (e.g. cod), mandatory measures implied that any vessel using gear with a mesh size of less than 100 mm (TR2) in Area 4.a in the North Sea must fish exclusively with any of the Highly Selective Gears (HSGs). Examples of these are the Gamrie Bay Trawl or Faithlie Cod Avoidance Panel. This made a significant portion of the fleet to switch to 100 mm mesh, as they can target both Nephrops and fish. This confirms anecdotal information suggesting that in recent years, vessels fishing in Fladen have become more dual purpose in the sense that more vessels are now targeting fish (using TR1 gears) and no longer solely dependent on Nephrops. This implies that these vessels have to buy both quota and days. Further general information on the fishery can be found in the Stock Annex.

### 4.5.3 ICES advice in 2015

## The ICES conclusions in 2015 in relation to State of the Stock were as follows:

"The stock size has declined from the highest observed value in 2008 and is just below the MSY Btrigger. The 2015 abundance estimate is the lowest of the time-series. The harvest rate has declined in recent years and remains well below FMSY. "

The ICES advice in 2015 (for 2016) (Single-stock exploitation boundaries) was as follows:

MSY approach
"ICES advises that when the MSY approach is applied, catches in 2016 (assuming zero discards) should be no more than 6856 tonnes. If instead discard rates continue at recent values (average of 2012-2014), and there is no change in assumed discard survival rate, this implies landings of no more than 6847 tonnes.

In order to ensure the stock in this FU is exploited sustainably, management should be implemented at the functional unit (FU) level. Should the catch in this FU be lower than advised, the difference should not be transferred to other FUs."

### 4.5.4 Management

Total Allowable Catch (TAC) management is at the ICES Subarea level. Most Nephrops vessels operate TR2 gear ( $>=70$ and $<100 \mathrm{~mm}$ ) and are subject to the effort regulations of the cod recovery plan. In Scotland the Conservation Credits scheme is in operation and various technical measures apply to Nephrops vessels (as described in section 3.4).

### 4.5.5 Assessment

## Approach in 2016

The assessment of Nephrops in 2016 is based on examining trends in the UWTV survey data (1992-2015) and utilising an extensive series of commercial fishery data and follows the process defined by the benchmark WG 2009. The assessment approach is further described in the stock annex.

The provision of advice in 2016 followed the process of 2015, and attempts to incorporate decisions taken at WKFRAME (2010) for the provision of MSY advice. The approach was developed based on inter-sessional work carried out by participants of the benchmark and involved collaboration between WGNSSK and WGCSE. The UWTV based assessments have derived predicted landings by applying a harvest rate approach to populations described in terms of length compositions from the trawl component of the fishery. Considerations for setting Harvest Ratios (HR) associated with proxies for FMSY for Nephrops are described in the WGNSSK 2010 report.

### 4.5.6 Data available

## Commercial catch and effort data

Landings from this fishery are predominantly reported from Scotland, with small contributions from Denmark and England, and are presented in Table 4.5.1 and Figure 4.5.1. Total international landings (as reported to the WG) in 2015 were 1786 tonnes (over 2000 tonnes lower than the 2014 total), consisting mostly of Scottish landings with only 12 tonnes landed by other countries.

In previous years, concerns were expressed over the reliability of the effort Figures provided for Scottish Nephrops trawlers; effort Figures were unrealistically low in some areas, particularly Fladen. Investigation of the issue revealed a problem in the MSS Marine Laboratory database, where only the effort expended in the first statistical rectangle visited by a vessel during a trip was being output. This did not affect landings. An extraction of days absent effort data by the Marine Scotland data unit in Edinburgh covering the four main trawl gears landing Nephrops into Scotland produced higher Figures which capture all the effort. At the present time, these revised data cover the period 2000 to 2015 and only annual summaries are available.

Trends in Scottish effort of Nephrops trawlers and LPUE are shown in Figure 4.5.1 and Table 4.5.2. From 2015, effort data for this stock is expressed both in days fishing and kW days (there are no major differences in effort trends between those different units). Effort has been relatively stable from 2002 to 2010 but fell markedly in the last five years because of poor fishing and part the fleet relocating to other areas. The spatial contraction of the fishery is further confirmed by the VMS distribution of otter trawlers fishing in Fladen (2010-2015) shown in Figure 4.5.8. In recent years a decreasing numbers of trips have been taking place in FU 7 and in 2015, the south of the ground was the area where most fishing took place. LPUE has gradually increased since 2000 to a peak of over $620 \mathrm{~kg} /$ day in 2009. It has fallen since then until 2013 followed by a slight increase in 2014 and a further decrease in 2015. Danish LPUE data (1991-2015) are presented in Table 4.5.3. Effort has generally decreased over the time whilst LPUE has gradually increased to a high in 2009 followed by a decreasing trend until 2013. In 2014-2015, the Danish LPUE showed a very slight increase but remains much lower than in previous years.

Males consistently make the largest contribution to the landings (Figure 4.5.2). This is likely to be due to the varying seasonal pattern in the fishery and associated relative catchability (due to different burrow emergence behaviour) of male and female Nephrops. This is confirmed by the quarterly landings as shown in Figure 4.5.2. In 20142015 the landings were much lower in the second quarter of the year, a period when females would be expected to be more available for capture. In 2013-2015 landings were larger in the third and fourth quarters. Figure 4.5 .7 shows the quarterly sex ratio by number from 2000. The seasonality of Nephrops emergency behaviour is apparent with males dominating catches, in particular during winter time (quarters 1 and 4). In quarters 2 and 3 , females become more active and are more available to the fishery, although in FU 7 (unlike FU 8 and 9) the sex ratio is less seasonal and closer to 50:50 all year round. In the last three years the male proportion in quarter 2 is much higher than previously. This may be related with sampling noise associated with the recent decrease in landings (and sampling opportunities) in that quarter. Sex ratio data does not seem to show an overall increase of female proportion in catches in the time series, except for the last two winters (2014 and 2015) where male percentage in catches decreased to less than $50 \%$. Increased female catchability has been associated with stocks which are in a poor state (females may remain more active as they have been unable to mate due to lack of males in the population). It is unclear if this is the case in FU 7 but sex ratio monitoring in catches will continue to inform on potential shifts in the balance of the population.

Discarding of undersized and unwanted Nephrops has occurred in this fishery, and quarterly discard sampling has been conducted on the Scottish Nephrops trawler fleet since 2000. The discarding rate average (2000-2015) is approximately $8 \%$ by number in this FU. From 2008 discard rates have dropped below the long term average and in the last five years the discard rates have been close to zero. In 2015 no discards have been recorded in the observer trips conducted. This reduction in discard rate appears to be due to a change in the discard pattern with lower numbers of small individuals being caught and could also signal reduced recruitment and a tendency towards the use of larger mesh gears (see below on length compositions).

It is likely that some Nephrops survive the discarding process. An estimate of $25 \%$ survival has been assumed in order to calculate dead removals (landings + dead discards) from the population.

## Intercatch

Scottish data for 2015 were successfully uploaded into Intercatch. National data coordinators for other countries also uploaded data to Intercatch ahead of the 2016 WG and output length compositions were obtained in formats suitable for running the assessment.

## Length compositions

Length compositions of landings and discards are obtained during monthly market sampling and quarterly on-board observer sampling respectively. Although assessments based on detailed catch data analysis are not presently possible for this species, examination of length compositions can provide a preliminary indication of exploitation effects.

Figure 4.5.3 shows a series of annual length frequency distributions for the period 2000 to 2015. Catch (removals) length compositions are shown for each sex with the mean catch and landings lengths shown in relation to MLS $(25 \mathrm{~mm})$ and 35 mm . In both sexes
the mean sizes have been generally stable over time except for the last five years where a noticeable shift in the length distribution and an increase in the mean size has been observed for males and to a lesser extent, females.

Figure 4.5 .1 and Table 4.5 .4 show the series of mean sizes of larger Nephrops ( $>35 \mathrm{~mm}$ ) in the landings. This parameter might be expected to reduce in size if overexploitation were taking place but there is no evidence of this. The mean size of smaller animals $(<35 \mathrm{~mm})$ in the catch is also fairly stable through time until 2010 when an increase is noticeable which may be associated with lower recruitments. In the last five years the landings mean size increased but discarding stopped, this may signal a period of reduced recruitment but could possibly reflect the increasing use of more selective gears. The discard rate in 2015 was estimated to be zero (as in 2011-2013). Quantitative information on trends in gear changes is not currently available. A further difficulty in the interpretation of these size observations is that the ground extends over a wide area and the distributional pattern of fleet activity is known to vary over time. This may lead to exploitation of sub-areas within the ground, where size compositions may be slightly different.

Mean weights in the landings through time (1990-2015) are shown in Figure 4.5.4 and Table 4.5.5. The variability in mean size is greater in Fladen (and Devil's Hole) than in other areas. In 2015 the mean weight in landings decreased to 36.7 g (following the mentioned reduction in the mean size of males).

## Natural mortality, maturity at age and other biological parameters

Biological parameter values are included in the Stock Annex.

## Research vessel data

Underwater TV (UWTV) surveys using a stratified random design are available for FU 7 since 1992 (missing survey in 1996). UWTV surveys of Nephrops burrow density and distribution reduces the problems associated with traditional trawl surveys that arise from variability in burrow emergence of Nephrops.

The numbers of valid stations used in the final analysis in each year are shown in Table 4.5.6. On average, approximately 65 stations have been considered valid each year ( 71 stations in 2015). Data are raised to a stock area of $28153 \mathrm{~km}^{2}$ based on the stratification (by sediment type). General analysis methods for UWTV survey data are similar for each of the Scottish surveys, and are described in more detail in the Stock Annex.

Previous review groups have noted that the UWTV survey did not cover the stock distribution. The survey stations are randomly distributed within strata and therefore the actual location of the survey stations varies from year to year and in some years, particular regions of the main part of the ground may not be surveyed. There is an additional small patch of mud to the north of the ground which it is not possible to survey (due to time constraints and distance to survey ground) and therefore the estimated absolute abundance is likely to be slightly underestimated by the UWTV survey.

### 4.5.7 Data analyses

## Exploratory analyses of survey data

Table 4.5.7 shows the basic analysis (corrected to absolute values) for the three most recent UWTV surveys conducted in FU 7. The table includes estimates of abundance and variability in each of the strata adopted in the stratified random approach. The
ground has a range of mud types from soft silty clays to coarser sandy muds ( $<40 \%$ silt and clay) and the latter predominates. Most of the variance in the survey is associated with the coarse sediment which surrounds the main centres of abundance.

Figure 4.5 .5 shows the distribution of stations in recent UWTV surveys (2010-2015), with the size of the symbol reflecting the Nephrops burrow density. Abundance is generally higher in the soft and intermediate sediments located to the centre and south east of the ground. Table 4.5.6 and Figure 4.5 .6 show the time series estimated abundance for the UWTV surveys, with $95 \%$ confidence intervals on annual estimates. Following the recent low UWTV estimated densities and the apparent Nephrops fleet preference for the fishing grounds located to the south of Fladen (Figure 4.5.8), the 2016 WG looked closely at the spatial distribution of the UWTV survey in the last seven years. It was suggested (as a hypothesis) that the north of the ground has been more affected by the recent decline (from 2009) in abundance than the areas in the south where most fishing took place in 2015. To test this, the TV surveys from 2009-15 were re-worked by sediment type, splitting the ground in two areas, north and south of the 58.75 N latitude line. Results seem to support that the areas mostly affected by the reduction in the mean Nephrops burrow density from 2009 were in fact located in the south, especially those made of finer sediments located in the central south region (Figure 4.5.9). In the north of Fladen, where coarser sediments ( $<40 \%$ silt and clay) dominate, a decrease in density was also observed but to a lesser extent when compared with those in the south. This analysis also shows that despite the recent decrease in density in the south, the mean densities remain in average higher than in the north.

The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009). A number of potential biases were highlighted including those due to edge effects, species burrow mis-identification and burrow occupancy. The cumulative bias correction factor estimated for FU7 was 1.35 meaning that the raw UWTV survey is likely to overestimate Nephrops abundance by $35 \%$. In order to convert the raw UWTV survey abundance to an absolute abundance the raw data are divided by 1.35 .

## Final assessment

The UWTV survey is again presented as the best available information on the Fladen Nephrops stock. This survey provides a fishery independent estimate of Nephrops abundance. At present it is not possible to extract any length or age structure information from the survey and it therefore only provides information on abundance over the area of the survey.

The 2015 UWTV survey data shows that the abundance has decreased $14 \%$ from the 2014 estimate. The stock remains at a low level and (below the average abundance over the time series) and is currently below the biomass trigger. The harvest ratio in 2015 ( $2 \%$, calculated as dead removals/TV abundance) is well below FMSY. The effort by Nephrops trawlers and respective LPUE declined from 2010 and this appears to be consistent with the abundance trends from the UWTV survey. The LPUE in recent years is still higher than the period prior to 2006 but this may be due the under-reporting of landings before the introduction of 'Buyers and Sellers' legislation. The relatively high LPUEs calculated for the period 2010-15, after the stock have declined could also be explained by the fishing fleet targeting areas where the density of Nephrops is higher. The mean size of individuals $>35 \mathrm{~mm}$ in the catch shows a clear increase. The discard rate remain at a very low level (average of $0.8 \%$ by number in 2013-2015) and the mean size of individuals below 35 mm shows an increasing trend from 2010, which may suggest a period of lower recruitment. Larger square mesh panels and new, more selective

TR2 gears implemented from 2010 as part of the Scottish Conservation Credits scheme in Division 4a may also have improved the exploitation pattern and reduced catches of smaller individuals.

## Historical Stock trends

The UWTV survey estimates of abundance for Nephrops in the Fladen suggest that the population has fluctuated over the 20 year period of the surveys. From 1997 to 2008 the abundance has generally increased and reached a peak of 7360 million individuals in 2008. Since 2008 the abundance has fallen until 2012 and increased slightly in the following two years. In 2015 the abundance has fallen again to 2569 million which is the lowest point of the time series (Table 4.5.8).

Table 4.5.8 also shows the estimated harvest ratios from 2003-2015. These range from $2-10 \%$ over this period and are all below $\mathrm{F}_{0.1}$. It is unlikely that prior to 2006, the estimated harvest ratios are representative of actual harvest ratios due to under-reporting of landings. In 2015, due to a $57 \%$ reduction in landings in relation to the previous year, the harvest ratio was estimated to be $2 \%$, which is also the lowest value recorded.

In addition to the discard rate, Table 4.5 .8 shows the dead discard rate which is the quantity of dead discards as a proportion (by number) of the removals (landings + dead discards).

### 4.5.8 Recruitment estimates

Recruitment estimates from surveys are not available for this FU. However, the increase in mean size of small animals $<35 \mathrm{~mm}$ (i.e. a lower proportion of small animals in this component of the catch) observed in recent years (2010-2015) may be indicative of lower recent recruitment.

### 4.5.9 MSY considerations

FMSY proxies for Nephrops are obtained from the per-recruit analysis as documented in the WGNSSK 2015 report. The most recent analysis used 2012-14 catch-at-length data, to account for the apparent changes in the discard pattern in this fishery. Length frequency data in Fladen have clearly shifted towards larger animals since 2010 (see section 4.5.5 and Figure 4.5.3) suggesting a different selection pattern in the fishery. In addition, the discard rate has declined (average of $7 \%$ by number in 2008-10 and around $0 \%$ in recent years), potentially due to a shift to larger meshes (TR1) and the increase in the use of the use of Highly Selective Gears for reducing fish bycatch. The biological parameters used in the analysis can be found in the Stock Annex. The complete range of the per-recruit FMSY proxies is given in the table below and the basis for choosing an appropriate FMSY proxy remains the same and is described in WGNSSK 2010 report.

| WGNSSK 2015 |  | Fbar(20-40 mм) |  | HR (\%) | SPR (\%) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | M | F |  | M | F | T |
| F0.1 | M | 0.07 | 0.07 | 6.4 | 47.4 | 58.3 | 51.9 |
|  | F | 0.14 | 0.15 | 10.6 | 33.3 | 40.8 | 36.4 |
|  | T | 0.08 | 0.09 | 7.5 | 43.0 | 53.1 | 47.2 |
| Fmax | M | 0.21 | 0.22 | 13.8 | 26.6 | 31.6 | 28.7 |
|  | F | 0.44 | 0.46 | 21.2 | 17.5 | 18.7 | 18.0 |
|  | T | 0.27 | 0.29 | 16.4 | 22.8 | 26.1 | 24.2 |
| F35\%SpR | M | 0.13 | 0.13 | 10.0 | 34.8 | 42.9 | 38.1 |
|  | F | 0.18 | 0.19 | 12.6 | 29.0 | 34.9 | 31.4 |
|  | T | 0.15 | 0.16 | 11.2 | 31.9 | 39.0 | 34.8 |

* $\mathbf{M}=$ males, $\mathrm{F}=$ females, $\mathrm{T}=$ combined

For this FU, the absolute density observed on the UWTV survey remains low (average of just below $0.2 \mathrm{~m}^{-2}$ ) suggesting the stock may have low productivity. In addition, the expansion of the fishery in this area is a relatively recent phenomenon and as a result the population has not been well-studied and biological parameters are considered particularly uncertain. Furthermore, historical harvest ratios in this FU have been below that equivalent to fishing at $\mathrm{F}_{0.1}$. For these reasons, it is suggested that a conservative proxy is chosen for FMSY such as $\mathrm{F}_{0.1(\mathrm{~T})}$.
The FMSY proxy harvest ratio is $7.5 \%$.
The $B_{\text {trigger }}$ point for this FU (lowest observed absolute UWTV abundance, 1992-2010) is calculated as 2767 million individuals.

### 4.5.10 Short-term forecasts

A catch prediction for 2017 was made for the Fladen Ground (FU7) using the approach agreed at the Benchmark Workshop in 2009 and outlined in the introductory section of the 2010 WGNSSK report. The table below shows catch predictions at various harvest ratios, including a selection of those equivalent to the per-recruit reference points discussed in Section 2 of this report and the harvest ratio in 2015 using the input parameters agreed at WKNEPH (ICES 2009). The catch prediction is calculated following the procedure outlined in the stock annex (section: short term projections).
Recently, to account for the landings obligation coming into force for Nephrops in 2016, the projected amount of discards (now referred to as unwanted catches) have been added to the catch options table. The advice given in 2016 considers three different scenarios: 1. Landings obligation applying for Nephrops with no discarding allowed; 2. Nephrops discarding is allowed to continue as before 2016; 3. Landings obligation with de minimis exemption applying for Nephrops with $6 \%$ discarding (by weight) under the minimum conservation reference size (MCRS, 25 mm in the North Sea) allowed. Three catch options tables are provided to account for each of these scenarios.

Under scenario 1, all catch is assumed to be landed, no discards will survive and therefore the harvest rate is assumed to include all catch and not only landings plus dead discards. Unwanted catches (by number) are calculated using data from the on-board observer sampling programme. This value is multiplied by the mean weight in discards to obtain the projected discard weight. A column is also included to show expected landings (referred to as wanted catches) under a landings obligation (Total Catches $=$ Wanted Catches + Unwanted Catches). The total catches calculated in this
way are lower than those calculated previously based on Landings + Surviving discards + Dead discards.

Under scenario 2, the catch options table includes surviving discards, in the same format as of the previous advice. This is to account for the possibility that Nephrops qualifies for a high survival exemption in which case a landings obligation would not be applicable. Discards survival for Nephrops in FU8 is assumed to be $25 \%$.

A de minimis exemption of $6 \%$ discards by weight below MCRS (Scenario 3) has been applied in the North Sea since the $1^{\text {st }}$ January 2016 and therefore, a catch options table accounting for a continuation of this rule in 2017 has been considered for the first time in the 2016 WG. The main difference from scenario 2 is that, under a de minimis exemption, if discard patterns remain unchanged, some unwanted animals above MCRS (these are typically soft animals with no commercial value) will have to be landed. As such, the catch options under this scenario include a new column for unwanted catch above MCRS (animals that would have been previously discarded) as this is not expected to be taken as landings. As all discarded animals are below MCRS under this assumption, the predicted weight of discards (dead + surviving) is lower than in scenario 2 ( $25 \%$ survival rate is still assumed).

The advice for Category 1 stocks (where assessment includes landings and discards data) is based on catches. The catch prediction for 2017 at the FMSY proxy harvest ratio under a de minimis exemption is 6844 tonnes. It should be noted that the FMSY proxy harvest ratio for Fladen is based on a combined Length Cohort Analysis (data 20122014) using dead removals (landings + dead discards). This value is expected to be updated in the future (using updated length information) to account for the landings obligation where no discard survival is assumed. A discussion of FMSY reference points for Nephrops is provided in Section 3.1.

The inputs to the landings forecast were as follows:

| Variable | Value | Source | Notes |
| :---: | :---: | :---: | :---: |
| Abundance in TV assessment | 2569 million | ICES (2016a) | UWTV 2015 |
| Mean weight in landings | 38.24 g | ICES (2016a) | Average 2013-2015 |
| Mean weight in discards | 15.30 g | ICES (2016a) | Average 2013-2015 |
| Mean weight in unwanted catch >MCRS | 16.13 g | ICES (2016a) | Average 2013-2015 |
| Mean weight in unwanted catch <MCRS | 7.58g | ICES (2016a) | Average 2013-2015 |
| Discard rate (total) | 0.83\% | ICES (2016a) | Average 2013-2015 (proportion by number) |
| Discard rate (>MCRS) | 0.67\% | ICES (2016a) | Average 2013-2015 (proportion by number) |
| Discard rate (<MCRS) | 0.16\% | ICES (2016a) | Average 2013-2015 (proportion by number) |
| Discard survival rate | 25\% | ICES (2016a) | Average 2013-2015 (proportion by number), only applies in scenarios where discarding is allowed. |
| Dead discard rate (total) | 0.62\% | ICES (2016a) | Average 2013-2015 (proportion by number), only applies in scenarios where discarding is allowed. |
| Dead discard rate (<MCRS) | 0.12\% | ICES (2016a) | Average (proportion by number) 20132015, only applies in scenarios where when discarding allowed for de minimus exemptions. |

Catch options assuming zero discards

|  | Total catch | Wanted Catch* | UnWanted <br> CATCH* | HARVEST RATE** |
| :--- | :--- | :--- | :--- | :--- |
| MSY approach | 6843 | 6821 | 22 | $7 \%$ |
| Fmsy | 7332 | 7308 | 24 | $7.5 \%$ |
| F2015 | 1955 | 1949 | 6 | $2 \%$ |
| F2013-2015 | 2835 | 2826 | 9 | $2.9 \%$ |
| F35\%SpR | 10949 | 10914 | 35 | $11.2 \%$ |
| Fmax | 16033 | 15981 | 52 | $16.4 \%$ |

[^0]Catch options assuming discarding is allowed

|  | Total <br> CATCH |  |  |  |  |  |  |  | Dead <br> REMOVALS | LANDINGS | DEAD <br> DISCARDS | SURVIVING <br> DISCARDS | HARVEST <br> RATE* |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | L+DD+SD | L+DD | L | DD | SD | for L+DD |  |  |  |  |  |  |  |
| MSY approach | 6858 | 6852 | 6835 | 17 | 6 | $7 \%$ |  |  |  |  |  |  |  |

* Calculated for dead removals.

Discarding allowed for de minimis excemptions only

| BASIS | Total CATCH | Dead REMOVALS | Landings | UnWANTED >MCRS* | Dead <br> DISCARDS <MCRS | Surviving DISCARDS | Harvest <br> RATE** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | L+DD+SD | L+DD | L | L | DD | SD | for $\mathrm{L}+\mathrm{DD}$ |
| MSY approach | 6844 | 6843 | 6822 | 19 | 2 | 1 | 7.0\% |
| Fmsy | 7334 | 7333 | 7310 | 21 | 2 | 1 | 7.5\% |
| F2015 | 1955 | 1955 | 1949 | 6 | 0 | 0 | 2.0\% |
| F2013-2015 | 2835 | 2835 | 2826 | 8 | 1 | 0 | 2.9\% |
| F35\%SpR | 10951 | 10950 | 10916 | 31 | 3 | 1 | 11.2\% |
| Fmax | 16035 | 16034 | 15984 | 46 | 4 | 1 | 16.4\% |

*Unwanted landings are those animals >MCRS but historically discarded
** Calculated for dead removals
$F_{0.1(\mathrm{~T})}$ : Harvest ratio equivalent to fishing at a level associated with $10 \%$ of the slope at the origin on the combined sex YPR curve.
$\mathrm{F}_{35 \% \mathrm{SPR}(\mathrm{T})}$ : Harvest ratio equivalent to fishing at a rate which results in combined SPR equal to $35 \%$ of the unfished level.
$F_{\max (T)}$ : Harvest ratio equivalent to fishing at a rate which maximises the combined YPR.

## Biological Reference points

Biological reference points have not been defined for this stock.

### 4.5.11 Quality of assessment

The TV surveys results show that the abundance has fallen in recent years, but not to the extent that would cause such a loss of fishing opportunity as observed. It is necessary to consider the biology of Nephrops (and indeed other crustaceans using cryptic, or burrow orientated behaviours) that are only available to trawling when they emerge from burrows. One explanation for the low emergence in 2013 (and to some extent in 2012) is that bottom temperatures appear to have been unusually low and for longer. Other environmental variables such as light levels, strength of tides are also known to exert an effect in the emergency behaviour of Nephrops. Exploratory analysis of the UWTV survey by sediment type (split by north and south of the ground) have shown that, despite the recent decrease in density, the mean densities remain in average higher in the south than in the north of FU 7 (see section 4.5.6). Taking into account the fact that the south of Fladen is located closer to the ports of Fraserburgh and Peterhead, where most of the fleet is based, this may explain why, in a period of lower densities, the south of FU 7 remains the area where most fishing activity takes place. Another factor that may play a role is that fishing in Fladen has become mixed in recent years and vessels may look for areas where economic returns are more favourable targeting both Nephrops and whitefish, while reducing fuel costs.

The recent low landings in Fladen may be the result of a complex interplay of factors combined with reduced average densities in the population as confirmed by the recent TV survey results.

The length and sex composition of the landings data is considered to be well sampled. Discard sampling has been conducted on a quarterly basis for Scottish Nephrops trawlers in this fishery since 2000, and is considered to represent the fishery adequately. Discard data covered $80 \%$ of the landings in 2015 (no discards were recorded).

The quality of landings (and catch) data is likely to have improved in recent years following the implementation of 'the registration of buyers and sellers' legislation in the UK in 2006, but because of concerns over the accuracy of earlier years, the final assessment adopted is independent of official statistics.

Underwater TV surveys have been conducted for this stock since 1992, with a continuous annual series available since 1997. The number of valid stations in the survey has remained relatively stable throughout the time period. Confidence intervals are relatively small.

The UWTV survey is conducted over the main part of the ground, representing an area of around $28200 \mathrm{~km}^{2}$ of suitable mud substrate (the largest ground in Europe). The Fladen Functional Unit contains several patches of mud to the north of the ground which are fished, bringing the overall area of substrate to $30633 \mathrm{~km}^{2}$. This area is not surveyed but would add to the abundance estimate. The absolute abundance estimate for this ground is therefore likely to be underestimated by the current methodology.

The Fishers' North Sea stock survey suggests that moderate or high amounts of recruits were apparent in Area 1 (which Fladen FU lies largely within) in 2011 compared to 2009. The time series of perceived abundance in Area 1 increases up to 2011. Opinion on discards appears to be split fairly evenly between lower, higher and no change. There are no Fishers' North Sea survey data available for 2013-2015.

### 4.5.12 Status of the stock

The stock has declined $63 \%$ in size in the period 2008-2012, then increased slightly in the 2013-2014 and in 2015 decreased again to the lowest point in the time series. The abundance is currently below the MSY $\mathrm{B}_{\text {trigger }}$ level. Landings taken from this FU in 2015 ( 1786 tonnes) were much lower than the 2014 advice (for 2015) of 10759 tonnes. The harvest rate decreased in 2015 to $2 \%$ and remains well below FMSY. Length frequencies in the caches have evolved towards larger animals over the last five years suggesting a selectivity change and/or lower recruitment.

### 4.5.13 Management considerations

The WG, ACOM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level. Management implemented at the Functional Unit level could provide controls to ensure that catch opportunities and effort were in line with the scale of the resource and that other FUs do not suffer from displacement from unused catch options from this FU.

Nephrops fisheries have a bycatch of cod. In 2005, high abundance of 0 group cod was recorded in Scottish surveys near to this ground. This year class of cod has subsequently contributed to slightly improved cod stock biomass and efforts are being made to avoid the capture of cod so that the stock can build further. The Scottish industry
operates under the Conservation Credits Scheme and is implementing improved selectivity measures in gears which target Nephrops and real time closures with a view to reducing unwanted by-catch of cod and other species.

### 4.5.14 References

ICES. 2012. Report of the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), 4-10 May 2011, ICES Headquarters, Copenhagen. ICES CM 2011/ACOM:13. 1197 pp.

ICES. 2015. Report of the Joint ICES-MYFISH Workshop to consider the basis for FMSY ranges for all stocks (WKMSYREF3), 17-21 November 2014, Charlottenlund, Denmark. ICES CM 2014/ACOM:64. 156 pp.

### 4.6 Firth of Forth (FU 8)

### 4.6.1 Ecosystem aspects

The Firth of Forth Functional Unit 8 is located in the south-west of the Northern North Sea and is an inshore ground just off the east coast of Scotland (Figure 3.1.1.). In common with other firths around the Scottish coast, the area is characterised by a wide entrance to seaward, narrowing towards the coast with river basins draining into the area. Sandy mud and muddy sand deposits are widespread throughout the area covering an area of $915 \mathrm{~km}^{2}$, the coarsest muds being found offshore beyond the May Island.

Owing to its burrowing behaviour, the distribution of Nephrops is restricted to areas of mud, sandy mud and muddy sand. Figure 4.6 .4 shows the distribution of sediment in the area. There is some evidence of Nephrops larval drift from grounds to the south of the area but most larvae appear to be produced locally and the population is characterised by high density and generally small size. Although this area was historically important for fish catches, this area has now declined and Nephrops is the main commercial species. The recruits of numerous demersal fish species occasionally aggregate in the area and small pelagics (sprat and juvenile herring) are seasonally abundant. Important seabird colonies occur in the area and the 'Wee Bankie' gravel area, important for sandeels is located further offshore to the north and east of the Firth.

### 4.6.2 The Fishery in 2015

The Nephrops fishery in the Firth of Forth is dominated by UK (Scotland) vessels with low landings reported by other UK nations (Table 4.6.1). In recent years around 40 vessels worked regularly in the Firth of Forth. Most vessels are under 12 m in length with about 10 in $12-15 \mathrm{~m}$ category and a few above 15 m . Engine power ranges from just under 100 kw to around the 300 kw . The trip length for most of the fleet is one day. In the winter, most vessels fish from around dawn till 16:00-19:00. In spring/summer, vessels switch to nights, working from around 19:00 to 07:00-10:00. The few larger vessels (over 15m) fishing in FU 8, undertake trips of around 2-3 days. The overall number of boats operating varies seasonally as vessels move around the UK in response to varying catch rates. In 2015 some large Fraserburgh boats, which usually operate in FU 7, moved into the area, fishing mostly to the east grounds of the Firth. Visitor boats come generally from the Northeast of Scotland (FU 7 and FU 9) in periods of poor fishing in those grounds. A few English vessels visited FU 8, mostly during summer, with landings from the rest of UK increasing from 22 tonnes in 2014 to 68 tonnes in 2015. Catches were generally reported as good with considerable market demand and a slight increase in prices for all sizes of Nephrops caught. Fuel prices have been reported as lower
than in previous years. The predominant trawl gear mesh sizes are 80 mm and 95 mm (TR2 gears with several vessels working with twin rigs). A few vessels have been involved with FDF (Fully documented Fisheries) getting some benefits in days at sea because of this. The fishery continues to be characterised by catches of small Nephrops which often leads to higher discard rates than in other east coast Functional Units. There was an increase in the amount of landings by creel vessels in this area to 43 tonnes in 2015 (14 tonnes in 2014) although typically the main target species of these vessels are crabs and lobsters.

Further general information on the fishery can be found in the Stock Annex.

### 4.6.3 Advice in 2015

## The ICES conclusions in 2015 in relation to State of the Stock were as follows:

"The stock size is above MSY Btrigger. The harvest rate increased in 2014 to 29.1\% and is now above FMSY."

The ICES advice in 2015 (for 2016) (Single-stock exploitation boundaries) was as follows:

MSY approach
"ICES advises that when the MSY approach is applied, catches in 2016 (assuming zero discards) should be no more than 2040 tonnes. If instead discard rates continue at recent values (average of 2012-2014), and there is no change in assumed discard survival rate, this implies landings of no more than 1866 tonnes.

In order to ensure the stock in this FU is exploited sustainably, management should be implemented at the functional unit level."

### 4.6.4 Management

Management is at the ICES Subarea level as described at the beginning of Section 4.2.

### 4.6.5 Assessment

## Approach in 2016

The assessment in 2016 is based on a combination of examining trends in fishery indicators and underwater TV using an extensive data series for the Firth of Forth Ground FU 8. The assessment of Nephrops through the use of the UWTV survey data and other commercial fishery data follows the process defined by the benchmark WG 2009 and described in the stock annex.

The provision of advice in 2016 followed the process of 2015, and attempts to incorporate decisions taken at WKFRAME (2010) for the provision of MSY advice. The approach was developed based on inter-sessional work carried out by participants of the benchmark and involving collaboration between WGNSSK and WGCSE. The UWTV based assessments have derived predicted landings by applying a harvest rate approach to populations described in terms of length compositions from the trawl component of the fishery. Considerations for setting Harvest Ratios (HR) associated with proxies for FMSY for Nephrops are described in the WGNSSK 2010 report.

## Data available

## Commercial catch and effort data

Landings from this fishery are predominantly reported from Scotland, with very small contributions from England, and are presented in Table 4.6.1 and Figure 4.6.1. Most of the landings are made by trawlers with creels accounting for just $2 \%$. Reported landings rose from 1100 to over 2650 tonnes between 2003 and 2009 and has fluctuated since then around 2000 tones. The value for 2009 of over 2,663 tonnes was the highest in the available time series whilst the 2015 landings (1892 tonnes) are below the ten year average (2200 tonnes).

In previous years, concerns were expressed over the reliability of the effort Figures provided for Scottish Nephrops trawlers; effort Figures were unrealistically low in some areas. Investigation of the issue revealed a problem in the MSS Marine Laboratory database, where only the effort expended in the first statistical rectangle visited by a vessel during a trip was being output. This did not affect landings. An extraction of days absent effort data by the Marine Scotland data unit in Edinburgh covering the 4 main trawl gears landing Nephrops into Scotland produced higher Figures which capture all the effort. At the present time, these revised data cover the period 2000 to the present and only annual summaries are available.

Trends in Scottish effort and LPUE are shown in Figure 4.6.1 and Table 4.6.2. From 2015, effort data for this stock is expressed both in days fishing and kW days (there are no major differences in effort trends between those different units). Effort has shown a gradual decline over the time period. Some of this is recently attributable to the EU effort management regime although, as part of the Scottish conservation credits scheme, Nephrops vessels have been eligible for effort 'buy-backs'. LPUE rose in the early 2000s and since 2006 has stabilised at a relatively high level.

Males consistently make the largest contribution to the landings by weight (Figure 4.6.2), although the sex ratio does vary and in 2011 more females in the catches moved the ratio closer to $1: 1$. This situation continued in 2012-2013. The proportion of females in the landings has increased in other years too (for example 2008). This may be due to the change in seasonal effort distribution with greatest effort in the $3^{\text {rd }}$ quarter in 2008 when females are likely to be more available to the fishery (compared with a more evenly distributed seasonal effort pattern in 2007 - Figure 4.6.2). Figure 4.6 .6 shows the quarterly sex ratio by number from 2000. The seasonality of Nephrops emergency behaviour is evident with males dominating catches during winter time. In quarters 2 and 3 females become more active and are more available to the fishery. These data suggest a gradual increase of female proportion in catches in recent years. Increased female catchability has also been associated with stocks which are in a poor state (females may remain more active as they have been unable to mate due to lack of males in the population). This problem usually manifests itself at times of the year when females would normally be reduced in the catches. This does not appear to be the case here.

Discarding of undersized and unwanted Nephrops occurs in this fishery, and quarterly discard sampling has been conducted on the Scottish Nephrops trawler fleet since 1990. Historically, discard rates have been higher in this stock than the more northerly North Sea FUs for which Scottish discard estimates are also available. This could arise from the fact that the use of larger meshed nets is not so prevalent in this fishery ( $80-95 \mathrm{~mm}$ is more common) and in addition, the population appears to consist of smaller individuals due to slower growth. Discarding rates in this FU have varied between 19\%
and $55 \%$ of the catch by number (2006-2015 average $31 \%$ ). In the last five years, discard rates appear to have dropped to below this value ( $25 \%$ on average by number). This appears to be due to increased retention of Nephrops rather than an absence of small Nephrops from the catches.

It is likely that some Nephrops survive the discarding process, an estimate of $25 \%$ survival is assumed in order to calculate dead removals (landings + dead discards) from the population.

## Intercatch

Scottish data for 2015 were successfully uploaded into Intercatch. National data coordinators for other countries also uploaded data to Intercatch ahead of the 2016 WG and output length compositions obtained in formats suitable for running the assessment.

## Length compositions

Length compositions of landings and discards are obtained during monthly market sampling and quarterly on-board observer sampling respectively. Although assessments based on detailed annual catch data analysis are not presently possible, examination of length compositions may provide an indication of exploitation effects.

Figure 4.6 .3 shows a series of annual length frequency distributions for the period 2000 to 2015. Size information on catches (removals) are shown for each sex with the mean catch and landings lengths shown in relation to MLS and 35 mm . There is little evidence of change in the mean size of either sex over time and examination of the tails of the distributions above 35 mm shows no evidence of reductions in relative numbers of larger animals.

The observation of relatively stable length compositions is further confirmed in the series of mean sizes of larger Nephrops ( $>35 \mathrm{~mm}$ ) in the landings shown in Figure 4.6.1 and Table 4.6.3. This parameter might be expected to reduce in size if overexploitation were taking place but over the last 20 years has in fact been quite stable. The mean size in the catch in the $<35 \mathrm{~mm}$ category (Figure 4.6.1) also shows no particular trend although it has risen slightly in the period 2009-2014, followed by a small decrease in 2015. The recent increase in the lower tail of discarded length frequencies (Figure 4.6.3), the decrease in the mean size of animals below 35 mm (Figure 4.6.1) and a slight increase in the discard rate in 2015 suggest possible a better recruitment in 2015. .

Mean weight in the landings is shown in Figure 4.5 .4 and Table 4.5.5 and this shows no systematic changes over the time series.

## Natural mortality, maturity at age and other biological parameters

Biological parameter values are included in the Stock Annex.

## Research vessel data

TV surveys using a stratified random design are available for FU 8 since 1993 (missing surveys in 1995 and 1997). Underwater television surveys of Nephrops burrow number and distribution, reduce the problems associated with traditional trawl surveys that arise from variability in burrow emergence of Nephrops.

The numbers of valid stations used in the final analysis in each year are shown in Table 4.6.4. On average, about 44 stations have been considered valid each year. In 2015, there were 51 valid stations. Abundance data are raised to a stock area of $915 \mathrm{~km}^{2}$. General
analysis methods for underwater TV survey data are similar for each of the Scottish surveys, and are described in the Stock Annex.

A further non-surveyed area of sediment (Lunan Bay) exists just north of the Firth of Forth FU. There is a small Nephrops fishery in this area (off Arbroath), but the area is only surveyed on an irregular basis and therefore is not included in any estimates of abundance. The WG wishes to emphasise that this area is out-with the Firth of Forth functional unit, is considered as part of the 'other' North Sea Nephrops area and hence not further considered in this section

## Data analyses

## Exploratory analyses of survey data

Table 4.6 .5 shows the basic analysis for the three most recent TV surveys conducted in FU 8. The table includes estimates of abundance and variability in each of the strata adopted in the stratified random approach. The ground is predominantly of coarser muddy sand. Depending on the year, high variance in the survey is associated with different strata and there is no clear distributional or sedimentary pattern in this area. Densities observed in this FU are typically higher than those of the more northerly FUs in the North Sea.

Figure 4.6 .4 shows the distribution of stations in TV surveys, with the size of the symbol reflecting the Nephrops burrow density. Abundance is currently higher towards the eastern and central parts of the ground and around the Isle of May. Table 4.6.4 and Figure 4.6 .5 show the time series of estimated abundance for the TV surveys, with $95 \%$ confidence intervals on annual estimates. The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009). A number of potential issues were highlighted including those arising from edge effects, species burrow mis-identification and burrow occupancy. To take account of these effects, a cumulative correction factor of 1.18 was estimated for FU 8 and this is applied to raw counts in order to derive the absolute abundance.

## Final assessment

The underwater TV survey is again presented as the best available information on the Firth of Forth Nephrops stock. This survey provides a fishery independent estimate of Nephrops abundance. At present it is not possible to extract any length or age structure information from the survey, and it therefore only provides information on abundance over the area of the survey.

The UWTV abundance was relatively high in the period 2003 to 2008 but has shown a decreasing trend in 2008-2012. In the last 3 years the stock has fluctuated around 600 million individuals. The stock is currently above the average abundance over the time series and remains above the biomass trigger. The calculated harvest ratio in 2015 (dead removals/TV abundance) decreased markedly and is just above FMSY. This is the result of an increase in stock abundance combined with a $20 \%$ decrease in landings in 2015. The mean size of individuals $>35 \mathrm{~mm}$ in the catch show no strong trend in recent years but the mean size of individuals below 35 mm has shown a slight increase from 2009. Larger square mesh panels and new, more selective TR2 gears implemented from 2010 as part of the Scottish Conservation Credits scheme may have improved the exploitation pattern. The effect of these changes are not however, as evident as those observed in FU 7 and length frequencies in recent years remain relatively stable in the Firth of Forth.

### 4.6.6 Historical Stock trends

The TV survey estimate of abundance for Nephrops in the Firth of Forth suggests that the population decreased between 1993 and 1998 and then began a steady increase up to 2003. Abundance is estimated to have fluctuated without trend in the years since then. The abundance estimates from 2003-2015 (the period over which the survey estimates have been revised) are shown in Table 4.6.6. The stock is currently estimated to consist of 664 million individuals.

Table 4.6.6 also shows the estimated harvest ratios over this period. These range from $12-29$ \% over this period, with the upper range being the estimated value for 2014 (estimated harvest ratios prior to 2006 may not be representative of actual harvest ratios due to under-reporting of landings before the introduction of 'Buyers and Sellers' legislation). The estimated harvest rate in 2015 is $16.8 \%$ which is just above the estimated value at $\mathrm{Fmsy}_{\text {( }} 16.3$ \%).

In addition to the discard rate, Table 4.6 .6 also shows the dead discard rate which is calculated as the quantity of dead discards as a proportion (by number) of the removals (landings + dead discards).

### 4.6.7 Recruitment estimates

Survey recruitment estimates are not available for this stock.

### 4.6.8 MSY considerations

A number of potential FMSY proxies were obtained from the per-recruit analysis for Nephrops as documented in the WGNSSK 2010 report. The most recent analysis (in 2011) used 2008-10 catch-at-length data, to account for the apparent changes in the discard pattern in this fishery. The biological parameters used in the analysis can be found in the Stock Annex. The complete range of the per-recruit FMSY proxies is given in the table below and the process for choosing an appropriate FMSY proxy is described in WGNSSK 2010 report.

| WGNSSK 2011 |  | Fbar(20-40 mм) |  | HR (\%) | SPR (\%) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | M | F |  | M | F | T |
| F0.1 | M | 0.14 | 0.06 | 7.7 | 40.8 | 62.3 | 49.9 |
|  | F | 0.31 | 0.13 | 15.2 | 20.5 | 40.7 | 29.0 |
|  | T | 0.17 | 0.07 | 9.4 | 34.6 | 56.6 | 43.9 |
| Fmax | M | 0.25 | 0.11 | 12.7 | 25.3 | 46.8 | 34.4 |
|  | F | 0.64 | 0.28 | 26.7 | 9.1 | 22.9 | 14.9 |
|  | T | 0.34 | 0.14 | 16.3 | 18.8 | 38.5 | 27.1 |
| F35\%SpR | M | 0.17 | 0.07 | 9.4 | 34.6 | 56.6 | 43.9 |
|  | F | 0.39 | 0.17 | 18.3 | 16.0 | 34.5 | 23.9 |
|  | T | 0.25 | 0.11 | 12.7 | 25.3 | 46.8 | 34.4 |

For this FU, the absolute density observed in the UWTV survey is relatively high (average of $\sim 0.7 \mathrm{~m}^{-2}$ ). Harvest ratios (which are likely to have been underestimated prior to 2006) have been well above $\mathrm{F}_{\max }$ and in addition there is a long time series of relatively stable landings (average reported landings ~ 2000 tonnes, well above those predicted by currently fishing at $\mathrm{F}_{\max }$ ) suggesting a productive stock. For these reasons, it is suggested that the sexes combined $\mathrm{F}_{\max (\mathrm{T})}$ is chosen as the FMSY proxy.

The FMSY proxy harvest ratio is 16.3 \%.

The Btrigger point for this FU (lowest observed absolute UWTV abundance) is calculated as 292 million individuals.

### 4.6.9 Short-term forecasts

A catch prediction for 2017 was made for the Firth of Forth (FU8) using the approach agreed at the Benchmark Workshop and outlined in the introductory section to this chapter (Section 3.1). The table below shows catch predictions at various harvest ratios, including a selection of those equivalent to the per-recruit reference points discussed in Section 2 of this report and the harvest ratio in 2015 using the input parameters agreed at WKNEPH (ICES 2009). The catch prediction is calculated following the procedure outlined in the stock annex (section: short term projections).

Recently, to account for the landings obligation coming into force for Nephrops in 2016, the projected amount of discards (now referred to as unwanted catches) have been added to the catch options table. The advice given in 2016 considers three different scenarios: 1. Landings obligation applying for Nephrops with no discarding allowed; 2. Nephrops discarding is allowed to continue as before 2016; 3. Landings obligation with de minimis exemption applying for Nephrops with $6 \%$ discarding (by weight) under the minimum conservation reference size (MCRS, 25 mm in the North Sea) allowed. Three catch options tables are provided to account for each of these scenarios.

Under scenario 1, all catch is assumed to be landed, no discards will survive and therefore the harvest rate is assumed to include all catch and not only landings plus dead discards. Unwanted catches (by number) are calculated using data from the on-board observer sampling programme. This value is multiplied by the mean weight in discards to obtain the projected discard weight. A column is also included to show expected landings (referred to as wanted catches) under a landings obligation (Total Catches = Wanted Catches + Unwanted Catches). The total catches calculated in this way are lower than those calculated previously based on Landings + Surviving discards + Dead discards.

Under scenario 2, the catch options table includes surviving discards, in the same format as of the previous advice. This is to account for the possibility that Nephrops qualifies for a high survival exemption in which case a landings obligation would not be applicable. Discards survival for Nephrops in FU8 is assumed to be $25 \%$.

A de minimis exemption of $6 \%$ discards by weight below MCRS (Scenario 3) has been applied in the North Sea since the $1^{\text {st }}$ January 2016 and therefore, a catch options table accounting for a continuation of this rule in 2017 has been considered for the first time in the 2016 WG . The main difference from scenario 2 is that, under a de minimis exemption, if discard patterns remain unchanged, some unwanted animals above MCRS (these are typically soft animals with no commercial value) will have to be landed. As such, the catch options under this scenario include a new column for unwanted catch above MCRS (animals that would have been previously discarded) as this is not expected to be taken as landings. As all discarded animals are below MCRS under this assumption, the predicted weight of discards (dead + surviving) is lower than in scenario 2 ( $25 \%$ survival rate is still assumed).

The advice for Category 1 stocks (where assessment includes landings and discards data) is based on catches. The catch prediction for 2017 at the FMSY proxy harvest ratio under a de minimis exemption is 2122 tonnes. It should be noted that the FMSY proxy harvest ratio in the Firth of Forth is still based on a combined Length Cohort Analysis (data 2008-2010) using dead removals (landings + dead discards). A discussion of FMSY reference points for Nephrops is provided in Section 3.1.

The inputs to the landings forecast were as follows:

| Variable | Value | Source | Notes |
| :---: | :---: | :---: | :---: |
| Abundance in TV assessment | 664 million | ICES (2016a) | UWTV 2015 |
| Mean weight in landings | 21.81g | ICES (2016a) | Average 2013-2015 |
| Mean weight in discards | 10.74 g | ICES (2016a) | Average 2013-2015 |
| Mean weight in unwanted catch >MCRS | 13.71 g | ICES (2016a) | Average 2013-2015 |
| Mean weight in unwanted catch <MCRS | 7.25 g | ICES (2016a) | Average 2013-2015 |
| Discard rate (total) | 24.9\% | ICES (2016a) | Average 2013-2015 (proportion by number) |
| Discard rate (>MCRS) | 13.3\% | ICES (2016a) | Average 2013-2015 (proportion by number) |
| Discard rate (<MCRS) | 11.6\% | ICES (2016a) | Average 2013-2015 (proportion by number) |
| Discard survival rate | 25\% | ICES (2016a) | Average 2013-2015 (proportion by number), only applies in scenarios where discarding is allowed. |
| Dead discard rate (total) | 19.9\% | ICES (2016a) | Average 2013-2015 (proportion by number), only applies in scenarios where discarding is allowed. |
| Dead discard rate (<MCRS) | 9.0\% | ICES (2016a) | Average (proportion by number) 20132015, only applies in scenarios where when discarding allowed for de minimus exemptions. |

Catch options assuming zero discards

| BASIS | Total CATCHES | WANTED <br> CATCHES* | UNWANTED <br> CATCHES* | HARVEST RATE** |
| :--- | :--- | :--- | :--- | :--- |
| MSY approach | 2062 | 1773 | 289 | $16.3 \%$ |
| F0.1 | 1189 | 1022 | 167 | $9.4 \%$ |
| F35SpR | 1607 | 1381 | 226 | $12.7 \%$ |
| F2015 | 2125 | 1827 | 298 | $16.8 \%$ |
| F2013_2015 | 2594 | 2230 | 364 | $20.5 \%$ |

* Wanted" and "unwanted" catch are used to described Nephrops that would be landed and discarded in the absence of the EU landing obligation based on discard rates estimates for average (2013-2015).
** Calculated for dead removals and applied to total catch.

Catch options assuming discarding is allowed

|  | Total CATCHES | Dead REMOVALS | LANDINGS | Dead DISCARDS | Surviving DISCARDS | Harvest RATE* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BASIS | L+DD+SD | L+DD | L | DD | SD | $\begin{aligned} & \text { for } \\ & \text { L+DD } \end{aligned}$ |


| MSY approach | 2199 | 2122 | 1890 | 232 | 77 | $16.3 \%$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

* Calculated for dead removals.

Discarding allowed for de minimis exemptions only

| BASIS | Total CATCH | Dead REMOVALS | Landings | UnWANTED >MCRS* | Dead DISCARDS <MCRS | Surviving DISCARDS | Harvest RATE** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | L+DD+SD | L+DD | L | L | DD | SD | $\begin{aligned} & \text { for } \\ & \text { L+DD } \end{aligned}$ |
| MSY approach | 2122 | 2099 | 1826 | 203 | 70 | 23 | 16.3\% |
| F0.1 | 1225 | 1211 | 1053 | 117 | 41 | 14 | 9.4\% |
| F35SpR | 1653 | 1635 | 1422 | 158 | 55 | 18 | 12.7\% |
| F2015 | 2187 | 2163 | 1882 | 209 | 72 | 24 | 16.8\% |
| F2013_2015 | 2669 | 2640 | 2296 | 256 | 88 | 29 | 20.5\% |

*Unwanted landings are those animals >MCRS but historically discarded
** Calculated for dead removals
$\mathrm{F}_{0.1(\mathrm{~T})}$ : Harvest ratio equivalent to fishing at a level associated with $10 \%$ of the slope at the origin on the combined sex YPR curve.
$\mathrm{F}_{35 \% \mathrm{SPR}(\mathrm{T})}$ : Harvest ratio equivalent to fishing at a rate which results in combined SPR equal to $35 \%$ of the unfished level.
$F_{\max (T)}$ : Harvest ratio equivalent to fishing at a rate which maximises the combined YPR.

## Biological Reference points

Biological reference points have not been defined for this stock.

### 4.6.10 Quality of assessment

The length and sex composition of the landings data is considered to be well sampled. Discard sampling has been conducted on a quarterly basis for Scottish Nephrops trawlers in this fishery since 1990, and is considered to represent the fishery adequately. Discard data covered $92 \%$ of the landings in 2015 ( $88 \%$ of the discards were imported and $12 \%$ were raised discards).

There are concerns over the accuracy of historical landings (pre 2006) due to misreporting and because of this the final assessment adopted is independent of officially reported data.

UWTV surveys have been conducted for this stock since 1993, with a continual annual series available since 1998.

The Fishers' North Sea Stock survey does not include specific information for the Firth of Forth. Area 3 shows a perception of decreased abundance over the period 2007 2012, but this covers the Firth of Forth and parts of the Devil's Hole in addition to the Moray Firth. There are no Fishers' North Sea survey data available for 2013-2015.

### 4.6.11 Status of the stock

The stock has declined in size since 2008 when it was at the highest point in the series but is still above the average abundance and well above the MSY Btrigger level. The UWTV abundance was relatively high in the period 2003 to 2010. The value calculated for 2015 ( 664 million) is about the same abundance to that recorded in 2013. Landings taken from this FU in 2015 (1892 tonnes) were higher than the 2014 advice (for 2015) of

1769 tonnes. The harvest rate decreased in 2015 to $16.8 \%$ and is just above Fmsy. Length frequencies in the catches have been stable.

### 4.6.12 Management considerations

The WG, ACOM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level. Management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort were compatible and in line with the scale of the resource.

Nephrops discard rates in this Functional Unit are relatively high in comparison to other Functional Units and there is a need to reduce these and to improve the exploitation pattern. An additional reason for suggesting improved selectivity in this area relates to bycatch. It is important that efforts are made to ensure that other fish are not taken as unwanted bycatch in this fishery which mainly uses 80 mm mesh. Larger square mesh panels and new, more selective TR2 gears implemented as part of the Scottish Conservation Credits scheme should help to improve the exploitation pattern for some species such as haddock and whiting and small cod.

Although the persistently high estimated harvest rates do not appear to have adversely affected the stock, they are estimated to be equivalent to fishing at a rate greater than FMSY and therefore it would be unwise to allow effort to increase in this FU.

### 4.7 Moray Firth (FU 9)

### 4.7.1 Ecosystem aspects

The Moray Firth Functional Unit is located in the east of the Northern North Sea and is an inshore ground just off the east coast of Scotland (Figure 3.1.1). In common with other firths around the Scottish coast, the area is characterised by a wide entrance to seaward, narrowing towards the coast with river basins draining into the area. Muddy sand deposits are the most widespread sediment, particularly towards the outer areas of the Firth, with smaller areas of sandy mud. Overall the ground covers an area of $2195 \mathrm{~km}^{2}$. In the inner parts of the Firth the sediment is more patchy and there are several areas of sand and of gravel.

Owing to its burrowing behaviour, the distribution of Nephrops is restricted to areas of mud, sandy mud and muddy sand. Figure 4.7 .4 shows the distribution of sediment in the area. It is thought that most larvae are produced locally although some drift from the Fladen may occur. The population is characterised by medium densities of Nephrops. Although the Moray Firth was historically important for whitefish fisheries, catches declined and Nephrops is the main commercial species with squid catches important in some years. The recruits of numerous demersal fish species occasionally aggregate in the area and small pelagics (sprat and juvenile herring) are seasonally abundant. The area is important for marine mammals (seals and cetaceans).

### 4.7.2 The Fishery in 2015

The Moray Firth Nephrops fishery is essentially a Scottish fishery with only occasional landings made by vessels from elsewhere in the UK (Table 4.7.1). Vessels targeting this fishery typically conduct day trips from the nearby ports along the Moray Firth coast. Around 15-20 local vessels (all single riggers) regularly fish in Moray Firth area, mostly out of Burghead. The majority of the Moray Firth fleet is under 10m and are not affected by Cod Recovery Measures. The fleet have been consistent in their grounds throughout
the years, with smaller vessels fishing locally from Burghead and larger and more powerful vessels venturing further out with good weather and reports of good catch rates in 2015. Occasionally larger vessels fish the outer Moray Firth grounds on their way to/from the Fladen or in times of poor weather. These larger (typically over 15 m ) twin riggers fished mainly in the outer areas of the Firth during the winter months and unlike the smaller local vessels, they can continue to operate in periods of poor weather. In 2012, a new voluntary code of conduct for Nephrops trawlers (Moray Firth Prawn Agreement) has been agreed amongst fishermen for the Inner Moray Firth so as to protect the viability of smaller vessels based in the area. The agreement proposes that an area in the most westerly part of the Moray Firth be reserved for vessels under 300HP with a further small area reserved for vessels under 400HP. Prices of Nephrops have been reported as similar to the previous years but fuel costs decreased considerably in 2015. Anecdotal evidence suggests some by-catch of monkfish and haddock occurred but vessels under 10 m , which make most of the fleet, are generally limited by quota restrictions. Nephrops creeling in the Moray Firth is not common (no landings in 2015) as grounds are in open water and gear conflicts with trawl vessels are likely to happen. A squid fishery usually takes place in the Moray Firth in the late summer, starting in the Southern Trench when squid moves inshore. The majority of the local Nephrops fleet participated in the squid fishery between September and October 2015. Further general information on the fishery can be found in the Stock Annex.

### 4.7.3 Advice in 2015

## The ICES conclusions in 2015 in relation to State of the Stock were as follows:

"The stock has declined in 2006 and has remained stable since then. The harvest rate has fluctuated around FMSY for the last decade"

The ICES advice in 2015 (for 2016) (Single-stock exploitation boundaries) was as follows:

MSY approach
"ICES advises that when the MSY approach is applied, catches in 2016 (assuming a landing obligation applies) should be no more than 943 tonnes. If this stock is not under the EU landing obligation in 2016 and discard rates do not change from the average (2012-2014), this implies landings of no more than 923 tonnes.

In order to ensure the stock in this FU is exploited sustainably, management should be implemented at the functional unit level."

### 4.7.4 Management

Management is at the ICES Subarea level as described at the beginning of Section 4.2.

### 4.7.5 Assessment

## Approach in 2016

The assessment in 2016 is based on a combination of examining trends in fishery indicators and UWTV using an extensive data series for the Moray Firth FU 9. The assessment of Nephrops through the use of the UWTV survey data and other commercial fishery data follows the process defined by the benchmark WG 2009 and described in the stock annex.

The provision of advice in 2016 followed the process of 2015, and attempts to incorporate decisions taken at WKFRAME (2010) for the provision of MSY advice. The approach was developed based on inter-sessional work carried out by participants of the benchmark and involved collaboration between WGNSSK and WGCSE. The UWTV based assessments have derived predicted landings by applying a harvest rate approach to populations described in terms of length compositions from the trawl component of the fishery. Considerations for setting Harvest Ratios (HR) associated with proxies for FMSY for Nephrops are described in the WGNSSK 2010 report.

## Data available

## Commercial catch and effort data

Landings from this fishery are predominantly reported from Scotland, with very small contributions from England, and are presented in Table 4.7.1. Total landings (as reported to the WG) in 2015 for Scotland were 828 tonnes (a $33 \%$ decrease in relation to 2014) and England landed 2 tonnes. Landings in recent years (post 2006) are more reliable due to the introduction of 'buyers and sellers' legislation. The long term landings trends are shown in Figure 4.7.1.

In previous years, concerns were expressed over the reliability of the effort Figures provided for Scottish Nephrops trawlers; effort Figures were unrealistically low in some areas. Investigation of the issue revealed a problem in the MSS Marine Laboratory database, where only the effort expended in the first statistical rectangle visited by a vessel during a trip was being output. This did not affect landings. An extraction of days absent effort data by the Marine Scotland data unit in Edinburgh covering the four main trawl gears landing Nephrops into Scotland produced higher Figures which capture all the effort. At the present time, these revised data cover the period 2000 to the present and only annual summaries are available.

Trends in Scottish effort and LPUE are shown in Figure 4.7.1 and Table 4.7.2. From 2015, effort data for this stock is expressed both in days fishing and kW days (there are no major differences in effort trends between those different units). Effort has shown a gradual decline over the time period. Some of this is attributable to the EU effort management regime although Nephrops vessels have generally been allocated exemptions. LPUE rose in the early 2000s and since 2006 it has fluctuated with a slightly downwards trend.

Males generally make the largest contribution to the landings by weight (Figure 4.7.2), although in 2011 the proportion of females is higher than in the recent past. In 20122015 males dominate again. The high contribution of females appears to be due to a much higher proportion of the fishery taking place in the second and third quarter when females are more available. This observation has been made a number of times before in the Moray Firth (particularly for example in 1994 when female catches exceeded those of males). Figure 4.7 .6 shows the quarterly sex ratio by number from 2000. The seasonality of Nephrops emergency behaviour is evident with males dominating catches during winter time. In quarters 2 and 3, females become more active and are more available to the fishery. These data suggest a fairly stable sex ratio in quarterly catches throughout the time series. Increased female catchability has also been associated with stocks which are in a poor state (females may remain more active as they have been unable to mate due to lack of males in the population). This problem usually manifests itself at times of the year when females would normally be reduced in the catches. This is not the case here.

Discarding of undersize and unwanted Nephrops occurs in this fishery, and quarterly discard sampling has been conducted on the Scottish Nephrops trawler fleet since 1990. Discarding rates in this FU appear to be highly variable with rates over the time series of 3 to $33 \%$ of the catch by number. In 2013 the observed rate by number was at its lowest level, approximately $3 \%$ by number but it increased to $15 \%$ in recent years (a similar level to that observed in 2010-2012). Discards rates were generally higher in the past and now appear to be generally lower but with occasional high annual levels which may be associated with occasional high recruitments (e.g. 2002 and 2004).

It is likely that some Nephrops survive the discarding process, an estimate of $25 \%$ survival is assumed in order to calculate dead removals (landings + dead discards) from the population.

## Intercatch

Scottish data for 2015 were successfully uploaded into Intercatch. National data coordinators for other countries also uploaded data to Intercatch ahead of the 2016 WG and output length compositions were obtained in formats suitable for running the assessment.

## Length compositions

Length compositions of landings and discards are obtained during monthly market sampling and quarterly on-board observer sampling respectively. Although assessments based on detailed catch analysis are not presently possible, examination of length compositions may provide an indication of exploitation effects.

Figure 4.7.3 shows a series of annual length frequency distributions for the period 2000 to 2015. Catch (removals) are shown for each sex with the mean catch and landings lengths shown in relation to MLS and 35 mm . There is little evidence of change in the mean size of either sex over time and examination of the tails of the distributions above 35 mm shows no evidence of reductions in relative numbers of larger animals. Occasional large year classes can be observed in these length frequency data (2002 and 2004). This is consistent with the occasional high discard rates observed for this FU.

The observation of relatively stable length compositions is further confirmed in the series of mean sizes of larger Nephrops ( $>35 \mathrm{~mm}$ ) in the landings shown in Figure 4.7.1 and Table 4.7.3. This parameter might be expected to reduce in size if overexploitation were taking place, but it appears to be stable throughout the time series. In 2013-2015, length frequencies seem to suggest a slight increase in the retention of larger males, which given the larger male contribution to the catches, caused an increase in the mean weight in the landings (Figure 4.5.4 and Table 4.5.5).

The mean size in the catch in the $<35 \mathrm{~mm}$ category (Figure 4.7.1) shows no particular trend over the time series although it has risen slightly in the period 2011-2013, followed by a small decrease in males in 2014 and both sexes in 2015, which is consistent with the recent (2014-2015) increase in the discard rate.

Natural mortality, maturity at age and other biological parameters
Biological parameter values are included in the Stock Annex.

## Research vessel data

Underwater TV (UWTV) surveys of Nephrops burrow number and distribution reduce the problems associated with traditional trawl surveys that arise from variability in burrow emergence of Nephrops.

The numbers of valid stations used in the final analysis in each year are shown in Table 4.7.4. On average, 41 stations have been considered valid each year, 52 stations were sampled in 2015. Abundance data are raised to a stock area of $2195 \mathrm{~km}^{2}$. General analysis methods for UWTV survey data are similar for each of the Scottish surveys, and are described in the Stock Annex.

## Data analyses

## Exploratory analyses of survey data

Table 4.7.5 shows the basic analysis for the three most recent UWTV surveys conducted in FU 9. The table includes estimates of abundance and variability in each of the strata adopted in the stratified random approach. The ground is predominantly of coarser muddy sand and typically, the variance in the survey seems to be evenly split among the different strata in recent years. The densities typically observed in this FU are lower than those observed in FU 8.

Figure 4.7.4 shows the distribution of stations in UWTV surveys, with the size of the symbol reflecting the Nephrops burrow density. In 2015, the abundance appears to be highest at the western end of the FU, with lower densities in the central and eastern areas. Table 4.7.4 and Figure 4.7.5 show the time series of estimated abundance for the UWTV surveys, with $95 \%$ confidence intervals on annual estimates. With the exception of 2003, the confidence intervals have been fairly stable in this survey.

The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009). A number of potential biases were highlighted including those due to edge effects, species burrow mis-identification and burrow occupancy. The cumulative bias correction factor estimated for FU 9 was 1.21 meaning that the TV survey is likely to overestimate Nephrops abundance by $21 \%$. In order to convert the raw UWTV survey abundance to an absolute abundance the raw data are divided by 1.21 .

## Final assessment

The UWTV survey is again presented as the best available information on the Moray Firth Nephrops stock. This survey provides a fishery independent estimate of Nephrops abundance. At present it is not possible to extract any length or age structure information from the survey and it therefore only provides information on abundance over the area of the survey.

The abundance in the Moray Firth has gradually decline since 2007 having increased in 2013 followed by a further decrease in 2014 and increased again slightly in 2015 (5\%) to 347 million. The stock is currently below the average abundance over the time series but remains above the biomass trigger. The calculated harvest ratio in 2015 (dead removals/TV abundance) is just below FMSY as a result of decreasing landings and a slight increase in stock abundance in 2015. The mean size of individuals $>35 \mathrm{~mm}$ in the catch shows no strong trend in recent years but the mean size of individuals below 35 mm has shown a slight increase from 2011. Larger square mesh panels and new, more selective TR2 gears implemented from 2010 as part of the Scottish Conservation Credits scheme may have improved the exploitation pattern as shown by a small increase in
the proportion of large males in caches in 2013-2015. The effect of these changes are not however, as evident as those observed in FU 7 and length frequencies in recent years remain relatively stable in the Moray Firth.

### 4.7.6 Historical Stock trends

The UWTV survey estimate of abundance for Nephrops in the Moray Firth suggests that the population increased between 1997 and 2005 but has gradually fallen since then (with the exception of 2007, 2013 and 2015). The abundance estimates from 2003-2015 are shown in Table 4.7.6.

Table 4.7.6 also shows the estimated harvest ratios over this period. These range from $7-20$ \% over the period 2003-2015. Estimated harvest ratios prior to 2006 may not be representative of actual harvest ratios due to under-reporting of landings before the introduction of 'Buyers and Sellers' legislation. The estimated harvest rate has decreased in 2015 to $9.1 \%$ and is now below the FMSY proxy value of $11.8 \%$.

In addition to the discard rate, Table 4.7.6 also shows the dead discard rate which is calculated as the quantity of dead discards as a proportion (by number) of the removals (landings + dead discards).

### 4.7.7 Recruitment estimates

Survey recruitment estimates are not available for this stock, although the length frequency distributions and highly variable discard rates suggest that this FU may be characterised by occasional large year classes.

### 4.7.8 MSY considerations

A number of potential FMSY proxies were obtained from the per-recruit analysis for Nephrops as documented in the WGNSSK 2010 report. The analysis was updated in 2011 using 2008-10 catch-at-length data, to account for the apparent changes in the discard pattern in this fishery and since previous estimates were derived several years before. An update was not performed this year. The complete range of the per-recruit FMSY proxies is given in the table below and the process for choosing an appropriate FMSY proxy is described in WGNSSK 2010 report.

|  |  | Fbar(20-40 mm) |  | HR (\%) | SPR (\%) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | M | F |  | M | F | T |
| F0.1 | M | 0.13 | 0.07 | 7.16 | 42.35 | 61.48 | 49.89 |
|  | F | 0.24 | 0.12 | 11.61 | 27.45 | 47.01 | 35.16 |
|  | T | 0.14 | 0.07 | 7.84 | 39.46 | 58.93 | 47.13 |
| Fmax | M | 0.26 | 0.13 | 12.31 | 25.80 | 45.16 | 33.42 |
|  | F | 0.68 | 0.36 | 23.82 | 11.42 | 25.16 | 16.83 |
|  | T | 0.34 | 0.18 | 14.92 | 20.79 | 39.10 | 28.01 |
| F35\%SpR | M | 0.17 | 0.09 | 9.11 | 34.69 | 54.48 | 42.48 |
|  | F | 0.41 | 0.22 | 17.12 | 17.62 | 34.83 | 24.40 |
|  | T | 0.24 | 0.13 | 11.79 | 27.02 | 46.53 | 34.71 |

The changes in the selection and discard patterns, and relative availability of females as estimated by the LCA result in slight decreases in the estimated MSY harvest ratio proxies compared to those calculated previously. (See stock annex for previously calculated values used at WGNSSK 2010).

Moderate absolute densities are generally observed on the UWTV survey of this FU (average of $\sim 0.2 \mathrm{~m}^{-2}$ ). Harvest ratios (which are likely to have been underestimated prior to 2006) appear to have been above $\mathrm{F}_{35 \% \text { SPR }}$ and in addition there is a long time series of relatively stable landings (average reported landings $\sim 1300$ tonnes, above those predicted by currently fishing at $\mathrm{F}_{35 \% \mathrm{SPR} \text { ). For these reasons, it is suggested that }}$ $\mathrm{F}_{35 \% \text { SPR(T) }}$ is used as the FMSY proxy.
The FMSY proxy harvest ratio is 11.8 \%.
The Btrigger point for this FU (lowest observed UWTV abundance) is calculated as 262 million individuals.

### 4.7.9 Short-term forecasts

A catch prediction for 2017 was made for the Moray Firth (FU9) using the approach agreed at the Benchmark Workshop. The table below shows catch predictions at various harvest ratios, including a selection of those equivalent to the per-recruit reference points discussed in Section 3.1 of this report and the harvest ratio in 2015 using the input parameters agreed at WKNEPH (ICES 2009). The catch prediction is calculated following the procedure outlined in the stock annex (section: short term projections).
Recently, to account for the landings obligation coming into force for Nephrops in 2016, the projected amount of discards (now referred to as unwanted catches) have been added to the catch options table. The advice given in 2016 considers three different scenarios: 1. Landings obligation applying for Nephrops with no discarding allowed; 2. Nephrops discarding is allowed to continue as before 2016; 3. Landings obligation with de minimis exemption applying for Nephrops with $6 \%$ discarding (by weight) under the minimum conservation reference size (MCRS, 25 mm in the North Sea) allowed. Three catch options tables are provided to account for each of these scenarios.

Under scenario 1, all catch is assumed to be landed, no discards will survive and therefore the harvest rate is assumed to include all catch and not only landings plus dead discards. Unwanted catches (by number) are calculated using data from the on-board observer sampling programme. This value is multiplied by the mean weight in discards to obtain the projected discard weight. A column is also included to show expected landings (referred to as wanted catches) under a landings obligation (Total Catches $=$ Wanted Catches + Unwanted Catches). The total catches calculated in this
way are lower than those calculated previously based on Landings + Surviving discards + Dead discards.

Under scenario 2, the catch options table includes surviving discards, in the same format as of the previous advice. This is to account for the possibility that Nephrops qualifies for a high survival exemption in which case a landings obligation would not be applicable. Discards survival for Nephrops in FU8 is assumed to be $25 \%$.
A de minimis exemption of $6 \%$ discards by weight below MCRS (Scenario 3) has been applied in the North Sea since the $1^{\text {st }}$ January 2016 and therefore, a catch options table accounting for a continuation of this rule in 2017 has been considered for the first time in the 2016 WG . The main difference from scenario 2 is that, under a de minimis exemption, if discard patterns remain unchanged, some unwanted animals above MCRS (these are typically soft animals with no commercial value) will have to be landed. As such, the catch options under this scenario include a new column for unwanted catch above MCRS (animals that would have been previously discarded) as this is not expected to be taken as landings. As all discarded animals are below MCRS under this assumption, the predicted weight of discards (dead + surviving) is lower than in scenario 2 ( $25 \%$ survival rate is still assumed).

The advice for Category 1 stocks (where assessment includes landings and discards data) is based on catches. The catch prediction for 2017 at the FMSY proxy harvest ratio under a de minimis exemption is 1070 tonnes. It should be noted that the FMSY proxy harvest ratio in the Moray Firth is still based on a combined Length Cohort Analysis (data 2008-2010) using dead removals (landings + dead discards. A discussion of FMSY reference points for Nephrops is provided in Section 3.1.

The inputs to the landings forecast were as follows:

| Variable | Value | Source | Notes |
| :---: | :---: | :---: | :---: |
| Abundance in TV assessment | 347 million | ICES (2016a) | UWTV 2015 |
| Mean weight in landings | 27.66 g | ICES (2016a) | Average 2013-2015 |
| Mean weight in discards | 11.45 g | ICES (2016a) | Average 2013-2015 |
| Mean weight in unwanted catch >MCRS | 14.37 g | ICES (2016a) | Average 2013-2015 |
| Mean weight in unwanted catch <MCRS | 6.63 g | ICES (2016a) | Average 2013-2015 |
| Discard rate (total) | 11.0\% | ICES (2016a) | Average 2013-2015 (proportion by number) |
| Discard rate (>MCRS) | 6.9\% | ICES (2016a) | Average 2013-2015 (proportion by number) |
| Discard rate (<MCRS) | 4.1\% | ICES (2016a) | Average 2013-2015 (proportion by number) |
| Discard survival rate | 25\% | ICES (2016a) | Average 2013-2015 (proportion by number), only applies in scenarios where discarding is allowed. |
| Dead discard rate (total) | 8.5\% | ICES (2016a) | Average 2013-2015 (proportion by number), only applies in scenarios where discarding is allowed. |
| Dead discard rate (<MCRS) | 3.1\% | ICES (2016a) | Average (proportion by number) 20132015 , only applies in scenarios where when discarding allowed for de minimus exemptions. |

Catch options assuming zero discards

| BASIS | Total <br> CATCH | WANTED CATCH* | UNWANTED CATCH** | HARVEST RATE** |
| :--- | :--- | :--- | :--- | :--- |
| MSY approach | 1060 | 1008 | 52 | $11.8 \%$ |
| F0.1 | 700 | 666 | 34 | $7.8 \%$ |
| F2015 | 817 | 777 | 40 | $9.1 \%$ |
| F2013_2015 | 889 | 846 | 43 | $9.9 \%$ |
| Fmax | 1338 | 1273 | 65 | $14.9 \%$ |

* Wanted" and "unwanted" catch are used to described Nephrops that would be landed and discarded in the absence of the EU landing obligation based on discard rates estimates for average (2013-2015).
** calculated for dead removals and applied to total catch.

Catch options assuming discarding is allowed

| BASIS | Total <br> Catches | Dead <br> Removals | Landings | Dead <br> DISCARDS | SURVIVING <br> DISCARDS | HARVEST <br> RATE* |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | L+DD+SD | L+DD | L | DD | SD | for <br> L+DD |
| MSY approach | 1089 | 1076 | 1036 | 40 | 13 | $11.8 \%$ |

* calculated for dead removals

Discarding allowed for de minimis excemptions only

| BASIS | Total CATCH | Dead REMOVALS | LANDINGS | UnWANTED >MCRS* | Dead <br> DISCARDS <br> <MCRS | Surviving DISCARDS | Harvest RATE** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | L+DD+SD | L+DD | L | L | DD | SD | for $\mathrm{L}+\mathrm{DD}$ |
| MSY approach | 1070 | 1067 | 1018 | 41 | 8 | 3 | 11.8\% |
| F0.1 | 708 | 706 | 673 | 27 | 6 | 2 | 7.8\% |
| F2015 | 826 | 824 | 785 | 32 | 7 | 2 | 9.1\% |
| F2013_2015 | 897 | 895 | 854 | 34 | 7 | 2 | 9.9\% |
| Fmax | 1353 | 1349 | 1286 | 52 | 11 | 4 | 14.9\% |

*Unwanted landings are those animals $>$ MCRS but historically discarded
** Calculated for dead removals
$\mathrm{F}_{0.1(\mathrm{~T})}$ : Harvest ratio equivalent to fishing at a level associated with $10 \%$ of the slope at the origin on the combined sex YPR curve.
$\mathrm{F}_{35 \% \mathrm{SPR}(\mathrm{T})}$ : Harvest ratio equivalent to fishing at a rate which results in combined SPR equal to $35 \%$ of the unfished level.
$F_{\max (T)}$ : Harvest ratio equivalent to fishing at a rate which maximises the combined YPR.

## Biological Reference points

Biological reference points have not been defined for this stock.

### 4.7.10 Quality of assessment

The length and sex composition of the landings data is considered to be well sampled. Discard sampling has been conducted on a quarterly basis for Scottish Nephrops trawlers in this fishery since 1990, and is considered to represent the fishery adequately. Discard data covered $44 \%$ of the landings in 2015 ( $50 \%$ of the discards were imported and $50 \%$ were raised discards). The reduction in the proportion of landings covered by discard data relates to missing sampling events in quarter 2 of the main metier (Nephrops trawlers, TR2 gears).

There are concerns over the accuracy of landings (pre 2006) and effort data and because of this the final assessment adopted is independent of official statistics.

UWTV surveys have been conducted for this stock since 1993, with a continual annual series available since 1996. The number of valid stations in the survey has remained relatively stable throughout the time period.

The Fishers' North Sea stock survey does not include specific information for the Moray Firth. Area 3 covers the Moray Firth, Firth of Forth and areas of the Devil's Hole and there appears to be some inconsistencies between the report in 2011 and 2012. In 2011 the report documented a perceived increase in the Nephrops abundance in this
area since 2008; however the 2012 report appears to show a perceived decrease since 2008. There are no Fishers' North Sea survey data available for 2013-2015.

### 4.7.11 Status of the stock

The evidence from the UWTV survey suggests that following a continuous decrease from 2007 to 2012 the abundance has fluctuacted around 400 million in recent years. The abundance has increased $5 \%$ in 2015 (to 347 million) remaining approximately at the same level as the late 2000's. The stock size is above the MSY Btrigger level. Landings taken from this FU in 2015 ( 830 tonnes) were lower than the 2014 advice (for 2015) of 1185 tonnes. The harvest rate decreased in 2014 to $9.1 \%$ and is now below $\mathrm{F}_{\text {MSY }}$ ( $11.8 \%$ ). Length frequencies in the catches have been relatively stable.

### 4.7.12 Management considerations

The WG, ACOM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level. Management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort were compatible and in line with the scale of the resource.
There is a by-catch of other species in the Moray Firth area. It is important that efforts are made to ensure that unwanted by-catch is kept to a minimum in this fishery. Current efforts to reduce discards and unwanted bycatches of cod under the Scottish Conservation credits scheme, include the implementation of larger meshed square mesh panels and real time closures to avoid cod.

The estimated harvest rates have generally been greater than FMSY and because the abundance (as estimated by the UWTV survey) appears to have declined in recent years, it would be unwise to allow effort to increase in this FU.

### 4.8 Noup (FU 10)

### 4.8.1 Ecosystem aspects

The Noup is a small area of muddy sand located to the west of Orkney. The area is exposed to the open Atlantic to the west and strong tidal currents occur in the area. The surrounding coarser grounds are important edible crab fishing areas and fish populations (mixed demersal species) are important in the locality.

### 4.8.2 The Fishery in 2014 and 2015

The Noup currently supports a relatively small fishery. Few vessels target Nephrops regularly in this area. In Orkney there is currently only one part-time (summer) vessel fishing for Nephrops as most of the local fleet targets crabs and lobsters. Nephrops boats from Orkney spend most of the year fishing in the Moray Firth (FU 9). In recent years, vessels from Scrabster landing Nephrops use 120 mm mesh twin rigs (targeting whitefish). Landings from Noup have decreased steadily since 2002 and in 2015 only 15 tonnes of Nephrops were landed (Table.4.8.1). Further general information on the fishery can be found in the Stock Annex.

### 4.8.3 Advice in 2014

The advice provided in 2014 was biennial and valid for 2015 and 2016.
"ICES advises on the basis of the ICES approach for data-limited stocks that catches should be no more than $33 t$. If discard rates do not change from the assumed rate of $7.9 \%$, this implies landings of no more than $32 t$.

To protect the stock in this functional unit (FU), management should be implemented at the functional unit level."

## Data available

## Commercial catch and effort data

Landings from this fishery are reported only from Scotland and are presented in Table 4.8.1 and Figure 4.8.1. Total landings (as reported to the WG) in 2015 were only 15 tonnes, the same as in the previous year and an increase of 2 tonnes from 2012 which was the lowest landing reported in the time series (1997-2015). Nephrops are almost exclusively landed by 'non-Nephrops' vessels. This supports the anecdotal information received from the fishing industry that this area is rarely fished by Nephrops vessels due to the high catch rates of whitefish in the area.

In previous years, concerns were expressed over the reliability of the effort Figures provided for Scottish Nephrops trawlers; effort Figures were unrealistically low in some areas. Investigation of the issue revealed a problem in the MSS Marine Laboratory database, where only the effort expended in the first statistical rectangle visited by a vessel during a trip was being output. This did not affect landings. An extraction of days absent effort data by the Marine Scotland data unit in Edinburgh covering the four main trawl gears landing Nephrops into Scotland produced higher Figures which capture all the effort. At the present time, these revised data cover the period 2000 to the present and only annual summaries are available.

Trends in Scottish effort and LPUE are shown in Figures 4.8.1 and Table 4.8.2. Effort on Nephrops trawlers (TR2) has declined over the time period and is more marked than on other Nephrops grounds owing to the dominance of trawlers targeting demersal fish in the area. LPUE remained approximately at the same level of 2014. From 2000 onwards landings, effort and LPUE have been split by TR1 and TR2 gears (Figure 4.8.2). Effort from TR1 in the last 10 years has remained relatively contstant but TR1 LPUE has declined from 2008 to 2010 remaining at a low level in the last 5 years.

## Length compositions

Levels of market sampling are low and discard sampling is not available. Mean sizes in the landings in previous years are shown in Figure 4.8.1 and Table 4.8.3. There are no sampling data available for 2015. The low levels of sampling for this fishery mean it is not realistic to draw conclusions from changes in size composition or sex ratio.

## InterCatch

Scottish data for 2015 were successfully uploaded into InterCatch prior the 2016 WG meeting according with the deadline proposed. Data for this stock is limited to official landings (classified as "Landing only" in Intercatch with no sampling data) provided by Scotland, therefore, length frequencies were not raised.

## Natural mortality, maturity at age and other biological parameters

No data available.

## Research vessel data

An underwater TV (UWTV) survey of this FU has been conducted sporadically (1994, 1999, 2006 and 2007). In 2014 Noup was re-visited by the summer Scotia UWTV survey after seven years past the previous survey. Figure 4.8 .3 shows the distribution of stations in the UWTV surveys, with the size of the symbol reflecting the Nephrops burrow density. In 2014, 12 stations were successfully surveyed. The most recent survey gives an estimate of population size ( 51 million) similar to that found in 2006 and 2007 which is slightly lower than the 1999 value. All of these are lower than the very high value observed in 1994. The results of the UWTV surveys are shown in Figure 4.8.4 and Table 4.8.4.

### 4.8.4 Historical stock trends

The TV survey estimate of abundance for Nephrops in the Noup suggests that the population declined from the first survey in 1994 to 1999 and remained at a lower level on the following surveyed years. Landings fluctuated between 200 and 400 tonnes between 1995 and 2002, and declined markedly from then. Recent landings for this FU have been low, approximately 16 tonnes in 2013 and 15 tonnes in 2014-2015.

### 4.8.5 Recruitment estimates

There are no recruitment estimates for this FU.

### 4.8.6 Short-term Forecasts

No short-term forecasts are presented for this FU.

### 4.8.7 Status of the stock

The current state of the stock is unknown.

### 4.8.8 Management considerations

The WG, ACOM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level. Management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort were compatible and in line with the scale of the resource.

The Noup area supports a mixed fishery in which Nephrops are taken mainly by demersal trawlers targeting fish. It is important that efforts are made to ensure that unwanted by-catch is kept to a minimum in this fishery. Current efforts to reduce discards and unwanted by-catches of cod under the Scottish Conservation credits scheme, include the implementation of larger meshed square mesh panels and real time closures to avoid cod.

The advice guidance and category classification for data-limited stocks (DLS) was addressed at WKLIFE2 (ICES, 2012). The methodology for DLS Nephrops stocks is further described in the 2013 Benchmark report (WKNEPH, 2013). Following the procedure outlined (section 3.1.2), the spatial extent of the Nephrops grounds were estimated (based on BGS sediment maps) to provide a likely envelope for the total abundance of Nephrops in FU 10 (see table below). UWTV survey information on the mean density of Nephrops ( 0.13 Nephrops $/ \mathrm{m}^{2}$ ), from the 2014 survey, was used together with discard percentages, and mean weights taken from FU 9 (Moray Firth). The same advice as provided in 2014 of 33 tonnes (catch) results in a harvest ratio of $2.4 \%$ which is below the range of harvest ratios observed for other North Sea functional units (7.5-16\%) and
therefore considered precautionary. Additional options including an increase in $20 \%$ in catches (uncertainty cap) and the medium term (10 year) average were included in the table. All the options (with the exception of the time series maximum landing value) result in a harvest ratio lower than $7.5 \%$, reflecting the low exploitation level in recent years in FU 10. The proposed advice (given in 2016) for 2017 and 2018 was that catches should be no more than 40 tonnes ( 2014 advice $+20 \%$ ). In line with the advice for other stocks, total catches, wanted catches and unwanted catches expected under the landing obligation policy were added to the table. A second catch options table allowing for discarding under the de minimis exemption ( $6 \%<\mathrm{MCRS}$ by weight for North Sea Nephrops) is also included below. This assumes the discard patterns do not change from average values previously observed, and the total catch is split in wanted catch, unwanted catch > MCRS, unwanted catch < MCRS and surviving discards. For data limited stocks the discard survival is assumed to be zero.

Catch options assuming zero discards

| Basis | Total <br> Catch | Wanted Catch | Unwanted Catch | Range of potential densities (Nephrops per $\mathrm{m}^{2}$ ) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 0.05 | 0.1 | 0.13 | 0.15 | 0.2 | 0.3 | 0.4 | 0.6 | 0.8 |
| Recent average (2013-2015) | 16 | 15 | 1 | 2.98\% | 1.49\% | 1.15\% | 0.99\% | 0.74\% | 0.50\% | 0.37\% | 0.25\% | 0.19\% |
| 2014 Advice - 20\% | 26 | 25 | 1 | 4.97\% | 2.48\% | 1.91\% | 1.66\% | 1.24\% | 0.83\% | 0.62\% | 0.41\% | 0.31\% |
| 2014 Advice | 33 | 31 | 2 | 6.16\% | 3.08\% | 2.37\% | 2.05\% | 1.54\% | 1.03\% | 0.77\% | 0.51\% | 0.38\% |
| 2014 Advice + 20\% | 40 | 38 | 2 | 7.55\% | 3.77\% | 2.90\% | 2.52\% | 1.89\% | 1.26\% | 0.94\% | 0.63\% | 0.47\% |
| Average (2006-2015) | 75 | 71 | 4 | 14.10\% | 7.05\% | 5.42\% | 4.70\% | 3.53\% | 2.35\% | 1.76\% | 1.18\% | 0.88\% |
|  | 102 | 97 | 5 | 19.27\% | 9.63\% | 7.41\% | 6.42\% | 4.82\% | 3.21\% | 2.41\% | 1.61\% | 1.20\% |
| Maximum | 519 | 494 | 25 | 98.13\% | 49.06\% | 37.74\% | 32.71\% | 24.53\% | 16.35\% | 12.27\% | 8.18\% | 6.13\% |

Catch options assuming discarding allowed for de minimis exemptions

| Basis | Total Catch | Wanted Catch | Unwanted <br> Catch <br> >MCRS | de minimis discards <MCRS | Surviving discards | Range of potential densities (Nephrops per m2) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | 0.05 | 0.1 | 0.13 | 0.15 | 0.2 | 0.3 | 0.4 | 0.6 | 0.8 |
| Recent average (2013-2015) | 16 | 15 | 1 | 0 | 0 | 2.98\% | 1.49\% | 1.15\% | 0.99\% | 0.74\% | 0.50\% | 0.37\% | 0.25\% | 0.19\% |
| 2014 Advice - 20\% | 26 | 25 | 1 | 0 | 0 | 4.97\% | 2.48\% | 1.91\% | 1.66\% | 1.24\% | 0.83\% | 0.62\% | 0.41\% | 0.31\% |
| 2014 Advice | 32 | 31 | 1 | 0 | 0 | 6.16\% | 3.08\% | 2.37\% | 2.05\% | 1.54\% | 1.03\% | 0.77\% | 0.51\% | 0.38\% |
| 2014 Advice + 20\% | 40 | 38 | 2 | 0 | 0 | 7.55\% | 3.77\% | 2.90\% | 2.52\% | 1.89\% | 1.26\% | 0.94\% | 0.63\% | 0.47\% |
| Average (2006-2015) | 75 | 71 | 3 | 1 | 0 | 14.10\% | 7.05\% | 5.42\% | 4.70\% | 3.53\% | 2.35\% | 1.76\% | 1.18\% | 0.88\% |
|  | 102 | 97 | 4 | 1 | 0 | 19.27\% | 9.63\% | 7.41\% | 6.42\% | 4.82\% | 3.21\% | 2.41\% | 1.61\% | 1.20\% |
| Maximum | 519 | 494 | 20 | 5 | 0 | 98.13\% | 49.06\% | 37.74\% | 32.71\% | 24.53\% | 16.35\% | 12.27\% | 8.18\% | 6.13\% |

Basis for the catch options.

| VARIABLE | VALUE | Source | Notes |
| :--- | :--- | :--- | :--- |
| Density in TV <br> assessment | 0.13 Nephrops <br> m2 | ICES (2016a) | UWTV 2014 |
| Mean weight in <br> landings | 27.66 g | ICES (2016a) | Average 2013-2015 (from FU 9) |
| Mean weight in <br> discards | 11.45 g | ICES (2016a) | Average 2013-2015 (from FU 9) |
| Mean weight in <br> unwanted catch <br> $>$ MCRS | 14.37 g | ICES (2016a) | Average 2013-2015 (from FU 9) |
| Mean weight in <br> unwanted catch <br> <MCRS | 6.63 g | ICES (2016a) | Average 2013-2015 (from FU 9) |
| Discard rate (total) | $11.0 \%$ | ICES (2016a) | Average 2013-2015 (from FU 9, <br> proportion by number) |
| Discard rate <br> (>MCRS) | ICES (2016a) | Average 2013-2015 (from FU 9) |  |
| Discard rate <br> (<MCRS) | ICES (2016a) | Average 2013-2015 (from FU 9) |  |
| Discard survival <br> rate | $0 \%$ | ICES (2016a) | Discard survival is assumed to be zero. |
| Surface area <br> estimate | 409 km 2 | ICES (2007) | Benchmark estimate WKNEPH (2007) |

### 4.9 Norwegian Deep (FU 32)

### 4.9.1 Ecosystem aspects.

See stock annex (section A.3).

### 4.9.2 Norwegian Deep (FU 32) fisheries

See stock annex (section A.2). Maps showing the annual geographic distribution of the Danish fishery in FU 32 were provided for the first time in 2015. Maps showing the annual geographic distribution of the Norwegian trawl fishery (vessels $\geq 15 \mathrm{~m}$ ) in FU 32 (2011-2015) were provided for the first time in 2016.

### 4.9.3 Advice in 2015

Advice for Nephrops was updated in 2015.
In 2015, ICES noted for this stock that:

- The perceptions of this stock (FU 32) are based on Danish landings and effort data as well as mean sizes (CL) in landings and discards.
- The new Danish LPUE index shows a stepwise declining trend from the mid-1990s until present. However, it is not possible to determine whether this decrease in LPUE is due to changes in management or whether the decrease to some extent also reflects stock changes.
- The recent Danish landings from the stock are very small, but are fished in a restricted area. The low LPUE in both 2013 and 2014 might therefore imply stock size changes in the southern part of FU 32 .
- Trends in mean size in landings and catches (discards from 2015 onwards), and overall size distribution in catches have for many years indicated that the Nephrops stock in FU 32 is not over-exploited. However, trends in mean size of landings in 2013 and 2014 are difficult to interpret as are the highly varying discard ratios during the last two years.
- The WG concludes that the available data give a non-conclusive perception of stock status. The average annual landings over the last ten years are 554 t (2005-2014), while the short-term average landings are 302 t (2010-2014). The biomass estimates indicate that harvest ratios for this stock have always been very low ( $\leq 1 \%$ ), even in years when landings were highest.


### 4.9.4 Management

An overview of the management of Nephrops in FU 32 is given in the stock annex (section A.2). The EU fisheries in FU 32 take place mainly in the Norwegian zone of the North Sea. The EU fisheries are managed by a separate TAC for this area. For 2008 the agreed TAC for EU vessels was $1300 t$, and for 2009-2012, 1200 t . In 2013, the TAC was reduced to $1000 t$, following the ICES advice, and it has remained at this level since. The EU quota of Nephrops in Norwegian waters (area 04-N) is mainly allocated to Denmark (app. 95\%) with a small fraction of app. $5 \%$ to UK. There is no quota restriction currently for the Norwegian fishery. It is not prohibited to discard Nephrops in FU 32.

### 4.9.5 Assessment

## Data available

Landings data for the 2016 assessment (all fleets in 2015) have been uploaded using InterCatch.

## Catch

Dutch landings from FU 32 were incorporated in the report for the first time in 2010, but it was only in 2006 that the Netherlands landed more than a ton of Nephrops from this FU. International landings from the Norwegian Deep increased from less than 20 t in the mid-1980s to 1190 t in 2001 (Table 4.9.1, Figure 4.9.1). Since then, landings have declined due to a reduction of Danish landings, and total landings in 2013 amounted to only 191 t, the lowest Figure since 1992. In 2014, total landings increased slightly, but decreased again in 2015 to 192 t . The decreased Danish landings can be explained by increasing fuel costs, fewer vessels, and Nephrops landings now occurring mainly as bycatch in mixed fisheries. Danish vessels used to take $80-90 \%$ of the total landings, but since 2008 this percentage has decreased. In 2015, Denmark landed $57 \%$ of the total landings. Norwegian landings decreased from 2009 to 2014, but increased in 2015 to 82 t.

Since 2003, the Danish at-sea-sampling programme has provided discard estimates (Table 4.9.1). Danish discards are low due to the legislated 120 mm mesh size. The Danish discard ratio (discard as percentage of catch) was high in 2004-2005 (21-24\%), but decreased to $10-17 \%$ in the years 2006-2012. In 2013, estimated Danish discards were 68 t , and the Danish discard ratio increased to $35 \%$, while in 2014 and 2015 estimated Danish discards were only 5 and 6 t , respectively, resulting in very low discard ratios of $3 \%$ and $5 \%$. The low discards the last two years may indicate low recruitment to the stock. It should be noted that the 2014- and 2015-discards in FUs 3-4 also were low. There are no Norwegian discard data, and Norwegian discards are assumed to be zero.

## Length composition

Figure 4.9 .1 was changed in the 2015 report. Previously, average sizes in catches and landings have been presented. In 2015 , this was changed to average sizes in landings $\geq$ 40 mm (MLS) and in discards < 40 mm . When considering recruitment mean size of individuals $<$ MLS is more informative than mean size in catches. The average size of Nephrops in Danish landings ( $\geq 40 \mathrm{~mm}$ ) showed a general increasing trend for both males and females in the period 2005-2012 (Figure 4.9.1). This increase coincides with a sharp decrease in landings and may imply a lower exploitation pressure. However, the mean size of both males and females in the Danish landings decreased sharply from 2012 to 2013. In 2014, the mean size of landed males jumped back to the high 2012level, and remained at this level in 2015. The average size of landed females, on the other hand, has remained at the low 2013-level both in 2014 and 2015. The mean size of discards ( $<40 \mathrm{~mm}$ ) has fluctuated without trend since 2002. In the 2014-report it was suggested that a possible explanation for the decreased mean size of large Nephrops could be that the declining Danish fishery in 2013 contracted into an area with small Nephrops, possibly in the southern part of FU 32. However, the maps of the annual distribution of the Danish fishery show that there was no change in the distribution from 2012 to 2013 (Figure 4.9.2). The Danish fishery has been located in the southernmost part of FU 32 since 2011. It is unclear why the mean size of landings has varied so much in recent years. The Norwegian time series of mean size in catches was removed from Figure 4.9.1 in 2015, but will be updated with additional length data (total lengths converted to CL) at the upcoming benchmark in 2016 and will be included in the report in 2017.

The length frequency distributions of the Danish catches from the years 2007, 2010, 2012 and 2014 had a greater proportion of large Nephrops compared with former years (Figure 4.9.3). The 2013 and 2015 length frequency distributions, on the other hand, had a relatively smaller proportion of large specimens. In general, there are few individuals below the MLS of 40 mm due to the legislated 120 mm mesh size. Size distributions of catches from Norwegian coast guard inspections of Danish and Norwegian trawlers (Figure 4.9.4), have not been updated since 2012 due to lack of CL data. Total lengths converted to CL will be included in the Figure at the upcoming benchmark. Danish and Norwegian length frequency distributions in catches have been compared for 2006, 2007 and 2012, years for which data exist from both countries (Figure 4.9.5). Trends are similar.

## Natural mortality, maturity at age and other biological parameters

No data are available at present. Data from the Norwegian shrimp survey covering FU 32 were considered by the 2013 benchmark (WKNEPH 2013) for estimation of maturity at length. However, annual catches in the survey are too small for estimation of annual maturity values. Possibilities for obtaining maturity data and length-weight relationships from the Danish at-sea-sampling programme are investigated.

## Catch, effort and research vessel data

Effort and LPUE Figures for the period 1989-2015 are available from Danish logbooks (Table 4.9.2, Figure 4.9.1). In 2013, the Danish effort index was changed to kW days (formerly fishing days) (see stock annex, section B.4), as kW days account for temporal differences in vessel size. Days at sea and fishing days are presented in addition to kW days (Table 4.9.2). The time series of fishing days (earlier years' effort index) was updated in 2013. In 2016, all efforts numbers back to 1987 changed slightly due to some minor adjustments to the métier codes for the whole time series. The LPUE values thus
also changed slightly but the trend remained the same. The Danish LPUE index based on kW days shows a decreasing trend (Figure 4.9.1). This is in contrast to the former LPUE index based on fishing days that fluctuated without trend around a mean level of approximately $200 \mathrm{~kg}^{\text {day }}{ }^{-1}$. As kW days is a more representative effort index, the new LPUE index is a more representative catch rate index. However, due to changes in the management regime, changes in the LPUE index do not necessarily imply stock size changes (see below).

In the beginning of the 1990s, vessel size increased in the Danish fleet fishing in FU 32. This increase, and more directed fisheries for Nephrops in areas with previously low exploitation levels are probably partly responsible for the observed increase in the Danish LPUE in those years (Table 4.9.2, Figure 4.9.1). The Norwegian mesh size legislation was changed in 2002 (see stock annex, section A.2). The introduction of the larger mesh size of 120 mm may explain the decrease in LPUE (catch rate) from 1999 to 2001 with a subsequent stabilizing at a lower level relative to the late 1990s. However, the lower LPUE may also reflect a stock decrease as Danish landings in 1999 increased to > 1000 t and remained at this level until 2006. In 2007, individual vessel quotas were introduced in the Danish fishery. This resulted in the vessels buying up a lot of fish quotas and shifting their effort to fin fish rather than Nephrops. To get good catches of Nephrops vessels need to target this species by fishing at dusk/dawn when the animals are out of their burrows, as opposed to fin fish fisheries where good catches can be obtained around the clock. This change in management coincided with a decreasing LPUE (2008-2009) and the onset of steadily falling Danish landings. From 2012 to 2013 the Danish LPUE decreased by approximately $40 \%$ and remained at this low level also in 2014 and 2015. As there has been no change in the distribution of the Danish fishery since 2011 the low LPUE in 2013-2015 might imply an increased fishing pressure in the southern part of FU 32. Environmental changes resulting in lower Nephrops densities cannot be ruled out. The likely low recruitment to the stock in 2014 and 2015 may imply continued low catch rates.

The Danish effort increased from 2004 to 2006, but showed a strong decline in 2007 and has since continued decreasing to 576 kW days in 2015, the lowest observed effort since 1990. It has not been possible to incorporate 'technological creep' in the evaluation of the effort data. However, the use of twin trawls has been widespread for many years.

The 2013 benchmark (WKNEPH 2013) analysed the Norwegian LPUE Figures from bottom and shrimp trawls. The data from bottom trawls and shrimp trawls prior to 2011 are considered unsuitable for LPUE analyses (see stock annex, section B.4). With the introduction of Norwegian electronic logbooks, compulsory for all vessels $\geq 15 \mathrm{~m}$ length in 2011, two new time series from respectively bottom and shrimp trawls (single and twin) have been established. These new indices will be considered for inclusion in the assessment at the upcoming 2016 benchmark. As a large portion of the Norwegian fleet landing Nephrops in FU 32 consists of vessels < 15 m , especially north of $60^{\circ} \mathrm{N}$, the Norwegian logbook data still only cover part of the fleet. The electronic logbook data show that the Norwegian large vessel trawl fishery for Nephrops in FU 32 declined from 2012 to 2013 (Figure 4.9.6). In 2013-2014, the fishery was confined to the southernmost part of the functional unit as well as an area just west of Stavanger, while in 2015 some trawling also took place along the western rim of the Norwegian Trench.

The Norwegian annual shrimp survey shows that Nephrops is distributed in areas deeper than 100 m in FU 32 (Figure 4.9.7). (Areas shallower than 100 m are not covered by the survey). In 2016, most trawl stations in FU 32 had catches of Nephrops of less
than 1 kg per trawled nm. No Nephrops was caught on the northernmost trawl stations. Catch rates were similar in FU 32 and FU 3 (Skagerrak).

## Data analysis

## Review of last year's assessment

"Technical comments
The stock annex contains text on two options for further work: the Nephrops benchmark in 2013 (WKNEPH 2013) suggested that the data from the shrimp survey should be investigated more closely, And two new time series from Norwegian bottom and shrimp trawls will be established from 2011 onwards. These new data have not been presented yet.

The stock annex does not explain how the calculations from the advice catch option table are performed. It is recommended to add this using a working example, so that the audit can be performed correctly.

It is unclear why a 10 years discard rate is used ( $15.9 \%$ according to the catch option table), given the changes in the Danish fishery. The discard rate in 2014 is $5 /(143+5)=3.3 \%$, which is very low compared to before. This should be mentioned and explained in the advice sheet?

It guess it is linked to how the years with missing discards are treated in the calculation, but the 2005-2014 average of catches from table Table 6.3.20.6 from the advice sheet (using "total" + Danish discards) is 636 and not 642 in my calculation.

## Conclusions

The advice is the average catch of the last ten years. It seems OK but the variability of discards ratio and the low value in 2014 merits discussion."

## Exploratory analysis of catch data

There was no age based analysis carried out

## Exploratory analysis of survey data

Catches of Nephrops from the annual Norwegian shrimp trawl survey (back to 1997) are small and variable. The benchmark in 2013 (WKNEPH 2013) suggested that although small, these catches should be explored with the aim of establishing a biomass index time series for Nephrops in FU 32. This is now work in progress and will be presented at the upcoming 2016 benchmark.

## Final assessment

No age based numerical assessment is presented for this stock. The state of the stock was judged on the basis of basic fishery data.

### 4.9.6 Historic stock trends

The increase in mean size in landings from 2006 to 2012 in females and from 2005 to 2012 in males could indicate a lower exploitation pressure as this increase coincided with decreasing landings. Mean sizes in landings in 2013-2015 are difficult to interpret. The introduction of a new effort index ( kW days) in 2013 resulted in a stepwise declining trend in the new LPUE index, from the mid-1990s until present.

### 4.9.7 Recruitment estimates

There are no recruitment estimates for this stock. Fluctuations in catches of small Nephrops are used as a proxy for recruitment.

### 4.9.8 Forecasts

There were no forecasts for this stock.

### 4.9.9 Biological reference points

No reference points are defined for this stock.

### 4.9.10 Quality of assessment

The data available for this stock remain limited.

### 4.9.11 Status of stock

The perceptions of this stock (FU 32) are based on Danish landings and effort data as well as mean sizes (CL) in landings and discards. A new Danish effort index which accounts for temporal differences in vessel size was introduced in 2013 ( kW days as opposed to former fishing days). The effect of technological creep on the effective effort of the fishery is still not known. The new Danish LPUE index shows a stepwise declining trend from the mid-1990s until present. However, it is not possible to determine whether this decrease in LPUE is due to changes in management or whether the decrease to some extent also reflects stock changes. The recent Danish landings from the stock are very small, but are fished in a restricted area. The low LPUE in 2013-2015 might therefore imply stock size changes in the southern part of FU 32. Trends in mean size in Danish landings and discards and overall size distribution in catches have for many years indicated that the Nephrops stock in FU 32 is not over-exploited. However, trends in mean size of landings in 2013-2015 are difficult to interpret. The low catches of small Nephrops during the last two years indicate low recruitment to the stock.

The WG concludes that the available data give a non-conclusive perception of stock status. The average annual landings over the last ten years are 464 t (2006-2015), while the short-term average landings are 259 t (2011-2015). The biomass estimates indicate that harvest ratios for this stock have always been very low ( $<1 \%$ ), even in years when landings were highest.

### 4.9.12 Management considerations

For 2006-2008 the agreed TAC for EU vessels was 1300 t . This decreased to 1200 t in 2009-2012 and 1000 t in 2013-2016. The WG notes that there is no TAC for the Norwegian vessels fishing in FU 32.

The Danish at-sea-sampling programme provided a satisfactory number of observer trips in 2015. However, quarters 1 and 2 were not sampled. Possibilities for obtaining biological data from the at-sea-sampling programme are explored. Norwegian sampling of catches by the Norwegian coast guard should be improved. Sample weights are not recorded, not allowing calculation of catches by length. Discard and landings components are not sampled separately and discards can therefore not be estimated.

ICES provide catch advice for FU 32. Advice is given for two scenarios: with and without a discard ban. Following the procedure outlined in the stock annex (section H) a table of harvest ratios (see table below) was calculated.

The biomass estimates imply very low harvest ratios in FU 32 ( $<1 \%$ ), even in former years with high landings ( $1000-1200 \mathrm{t}$ ). ICES advise that when the precautionary approach is applied, catches in 2016 (assuming a landing obligation applies) should be no more than 496 t .

|  |  |  |  |  |  |  |  |  | 2015- <br> values |  |  | $\begin{aligned} & \text { mean (2006- } \\ & 2015) \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | FU32 : Norwegian Deep |  |  |  |  |  | 55500 | Area (km2 | 101 | land wt |  |  | percentage discards |
|  |  |  |  |  |  |  |  |  | 25 | disc wt |  |  |  |
| Landing obligation |  |  |  |  | Density |  |  |  |  |  |  |  |  |
| Basis | Total catch | Wanted catch | Unwanted catch |  | 0.05 | 0.1 | 0.2 | 0.3 | 0.4 | 0.5 | 0.6 | 0.7 | 0.8 |
| 0.5 * Average | 248 | 232 | 16 |  | 0.1\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Average | 496 | 464 | 32 |  | 0.2\% | 0.1\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Maximum | 1273 | 1190 | 83 |  | 0.5\% | 0.3\% | 0.1\% | 0.1\% | 0.1\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% |
|  |  |  |  |  |  | Fladen (FU7) density |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Discarding allo | owed |  |  |  |  |  |  |  |  |  |  |  |  |
| Basis | Dead removals | Landings | Dead discards | Live discards | 0.05 | 0.1 | 0.2 | 0.3 | 0.4 | 0.5 | 0.6 | 0.7 | 0.8 |
| 0.5 * Average | 244 | 232 | 12 | 4 | 0.1\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Average | 488 | 464 | 24 | 8 | 0.2\% | 0.1\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Maximum | 1252 | 1190 | 62 | 21 | 0.5\% | 0.3\% | 0.1\% | 0.1\% | 0.1\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% |
|  |  |  |  |  |  | Fladen (FU7) density |  |  |  |  |  |  |  |

### 4.10 Off Horns Reef (FU 33)

### 4.10.1 Data available

## Catch

The landings from FU 33 were marginal for many years. However, from 1997 to 2004, Danish landings increased considerably, from 274 to $1,097 \mathrm{t}$. Denmark dominated the fishery during this period. Since 2004, Danish landings have gradually decreased and in 2015 were 371 t . The other countries reporting landings from the area are Belgium, Netherlands, Germany and the UK. Dutch landings show an increasing trend from the start of the time series until 2007 when landings were almost 500 t . Since 2007, Dutch landings show a decreasing trend and in 2015 were the lowest landings recorded over the last decade ( 187 t ). Belgium and German landings having increased throughout the time period and were around 300 and 140 t respectively in 2015. UK landings were highest in 2009 ( 170 t ) and have since decreased dramatically. In 2007, total landings were the highest on record ( 1467 t ). Since 2007, total landings have remained relatively stable and fluctuated without trend, and were 1003 t in 2015 (Table 4.10.1 and Figure 4.10.1). Discard weights are only available for a portion of the fishery (Denmark and the Netherlands) and have been used to raise discards.

## Length compositions

Length (CL) distributions of the Danish catches 2001 to 2005 and 2009 to 2015 are shown in Figure 4.10.2. Notice, that except for 2005 and 2011 they are rather similar. Figure 4.10 .1 shows the development of the mean size of Nephrops in catches. The drop in the mean CL in the catches in 2005 and 2013 reflects an increase in numbers at around 30 mm CL and could indicate a large recruitment in these years, see also Figure 4.10.1

In the period 2001-2005, and in 2009-2015 the Danish at-sea-sampling programme has provided data for discard estimates. However, the samples do not cover all quarters.

## Natural mortality, maturity at age and other biological parameters

No data available

## Catch and effort data

Table 4.10.2 and Figure 4.10 .1 show the development in Danish effort and LPUE. Notice that the 10-fold increase in fishing effort from 1996 to 2004 seems to correspond to the increase in landings during the same period and the LPUE was relatively stable. After 2004 the Danish effort decreased markedly, and since 2009 has remained stable at around 300000 kW days. Dutch effort data are available from 2005-2015 and shows an increasing trend over the time period. However, Dutch effort decreased from around 1300000 kW days in 2013 to 1000000 kW days in 2014 and 2015. The Danish LPUE shows an increasing trend during the whole period, and since 2011 has remained above $1.0 \mathrm{~kg} / \mathrm{kW}$ day). This increase in LPUE could reflect an increase in gear efficiency (technological creep) or in fishers' ability to exploit the stock. LPUE from the Netherlands increased from $0.3 \mathrm{~kg} / \mathrm{kW}$ day in 2005 to around $0.7 \mathrm{~kg} / \mathrm{kW}$ day in 2007, and has since fluctuated between 0.2 and $0.5 \mathrm{~kg} / \mathrm{kW}$ day.

## Data analysis

## Exploratory analyses of catch data

No catch at age analysis has been carried out for this stock.

## Exploratory analyses of survey

No survey data were available

### 4.10.2 Historic stock trends

The available data do not provide any clear signals on stock development:
Danish effort began decreasing after 2004. Since then, the LPUE has steadily increased, except for 2010 when LPUE declined slightly. In 2013, new data from the Netherlands became available for the last nine years, and shows a more stable effort. LPUE has increased for Denmark and decreased for the Netherlands in 2015.

In 2015, the size distribution in the catches is similar to those in 2001-04, 2009-2010 and 2012-2013. The smaller individuals in the 2005 and 2011 catches could reflect a high recruitment in these years. The decrease in mean size could indicate either high recruitment or a decline in the stock, reflected by fewer large individuals. However, there are no recruitment estimates for this FU.

Forecasts: Forecasts were not performed.
Biological reference points: There are no reference points defined for this stock.
Perceptions of the stock are based on Danish and Dutch LPUE data and trends in size composition in Danish catches. As stated above, comparing the size distribution in the 2005 and 2011 catches with those in other years could indicate high recruitment in 2005 and 2011.

### 4.10.3 Management considerations for FU 33.

The North Sea TAC is not thought to be restrictive for the fleets exploiting this stock. Considering the recent trend in LPUE and the technological creep of the gear, the exploitation of this stock should be monitored closely.

### 4.10.4 Status of the stock

The state of this stock is unknown. Based on the assumed low density (based on lowest observed density at FU 7 (Fladen Ground), harvest rates are considered low for this stock.

The size of the Off Horns Reef Nephrops grounds are 5700 km 2 and are believed to have a density similar to the Fladen Grounds (FU 7) of 0.1 Nephrops $m-2$. The mean individual weight in landings and discards are 40.57 and 17.19 g respectively and the survival rate of discards is $25 \%$. Discards are known to take place for the entire fishery, however only length measured discard data exist for the Danish fishery. These data are believed to be representative for the entire fishery and have been used to calculate the values in the catch options table.

### 4.11 Devil's Hole (FU 34)

The Devil's Hole was designated as a functional unit in 2010, after recommendation from SGNEPS because of increasing landings in the area. The latest advice for this functional unit was provided in 2014 using the data limited approach.

### 4.11.1 Ecosystem aspects

The area consists of a number of narrow trenches (up to 2 km wide) running in a northsouth direction, with an average length of $20-30 \mathrm{~km}$. These trenches fall across six ICES statistical rectangles: 41-43F0 and 41-43F1, which are used to define this functional unit. The British Geological Survey (BGS) sediment map (showing sediments suitable for Nephrops) of the area is shown in Figure 4.11 .1 and suggests that there is one large, and several smaller areas of muddy sand ( $10-50 \%$ silt and clay).

### 4.11.2 The Fishery in 2014 and 2015

The fishery in this area is prosecuted largely by Scottish vessels operating out of ports in the northeast of Scotland, but occasionally making landings into northeast England. The fleet consists of large Nephrops trawlers which have the capability of operating in such offshore areas. Around five vessels operate out of Peterhead with another 12 from Fraserburgh regularly visiting the areas. These vessels also fish the Fladen on a regular basis and visit the other more inshore functional units in times of poor weather or poor Nephrops catch rates in the offshore areas. Effort in FU 34 by vessels with Nephrops gears has decreased from 2009 and fewer boats were reported to be operating in the Nephrops offshore grounds. In contrast, landings and effort from TR1 vessels have increased in recent years.

## Advice in 2014

Advice provided in 2014 was biennial for 2015 and 2016.
"ICES advises on the basis of ICES approach to data-limited stocks that catches should be no more than $410 t$. If discard rates do not change from the recent average (2008-2011), this implies landings of no more than 383 t .

To protect the stock in this functional unit (FU), management should be implemented at the functional unit level."

### 4.11.3 Management

Total Allowable Catch (TAC) management is at the ICES Subarea level.

### 4.11.4 Assessment

Data are presented which in future may form the basis for an assessment. A benchmark was carried out for this functional unit in 2013 (WKNEPH 2013) which advised to continue with the data limited approach at present with the aim of moving to a full underwater TV (UWTV) assessment in the near future.

### 4.11.5 Data available

## Commercial catch and effort data

Overall landings from this fishery for 1986-2015 are presented in Table 4.11.1 and Figure 4.11.2. Landings gradually increased from 378 tonnes in 2005 to approximately 1305 tonnes in 2009 followed by a decline in the following years to 121 tonnes in 2013. In 2015 landings increased to 439 tonnes. This is explained by an increase in both landings from TR1 (which target mostly whitefish) and TR2 vessels.

In previous years, concerns were expressed over the reliability of the effort Figures provided for Scottish Nephrops trawlers; effort Figures were unrealistically low in some areas. Investigation of the issue revealed a problem in the MSS Marine Laboratory database, where only the effort expended in the first statistical rectangle visited by a vessel during a trip was being output. This did not affect landings. An extraction of days absent effort data by the Marine Scotland data unit in Edinburgh covering the four main trawl gears landing Nephrops into Scotland produced higher Figures which capture all the effort.

Trends in Scottish effort and LPUE are shown in Figure 4.11.3 and Table 4.11.2. TR2 effort has fluctuated over the time period showing generally a downwards trend from 2003, reaching its lowest point in 2013. The decrease may partly be explained as a result of reductions in available effort imposed by the effort management regime and partly because this ground is more remote than a number of other Nephrops grounds and costs of steaming to and from the ground are likely to be high. In 2014-2015 TR2 effort increased again to a similar level to that recorded in 2012. TR1 effort decreased sharply in the early 2000's and has fluctuated ever since.

LPUE for Scottish Nephrops trawlers (TR2) showed an increasing trend from 2003 to 2009 followed by a slight drop in 2011 and has remained relatively stable in the last four of years. LPUE for TR1 trawlers has been fluctuating without trend over the time series.

## Length compositions

Levels of both market and discard sampling are low and data are only available from the Scottish fleet. Most observer sampling in FU 34 took place in the period 2008-2011. No market samples were taken in 2012-2013 and in the last two years only three fishing trips were sampled. Mean sizes in the catch and landings for 2006 to 2011 are shown Table 4.11.3. Sampling has not been conducted in all quarters, so there is potential bias in these results.

## InterCatch

Scottish data for 2015 were successfully uploaded into InterCatch prior the 2016 WG meeting according with the deadline proposed. Both landings and discard sampling have been very limited in recent years and Intercatch was used only to record official
landings data (no raising) from counties who submitted data into FU 34 (Scotland, England and Denmark).

## Natural mortality, maturity at age and other biological parameters

No specific data are available for this functional unit, but there may be potential to adapt parameters from other functional units which have apparently similar biological characteristics.

## Research vessel data

Marine Scotland Science (MSS) have carried out UWTV surveys of the Devil's Hole area opportunistically over the past 10 years. Since 2009, VMS data have been used to define the location of the survey stations. It is not known how station locations were selected on the earlier surveys in this area. It was not possible to survey FU 34 in 2013 but the survey has continued in recent years. The TV surveys in the period 2009-2014 were re-worked to exclude any station lying outside the VMS strata that was adopted in the 2013 benchmark (WKNEPH, 2013). Overall, 6 stations were excluded in the 5 year period and this had little impact on the new abundance values which are very similar to previous estimates. In 2015, FU 34 was visited by the summer Scotia UWTV survey and 17 stations were successfully surveyed. A decline was observed in the Nephrops density from 2009 to 2012 followed by a slight increase in recent years. The most recent survey give estimates of density ( 0.16 burrows $/ \mathrm{m}^{2}$ ) similar to those found in 2011. A density distribution map of these surveys is shown in Figure 4.11 .4 with the size of the symbol reflecting the Nephrops burrow density. Table 4.11.4 and Figure 4.11.5 show the time series of mean burrow densities and $95 \%$ confidence intervals.

### 4.11.6 Historical stock trends

Scottish landings from this area have risen substantially from 2005 to 2009 followed by a general decreasing trend until 2013 and increased again in the last 2 years. Estimates of mean density in the stock have declined from 2009 to 2014 but increased slightly in 2015, remaining higher than in 2003, although this may be due to the change is survey sampling design, with a greater proportion of stations in the western trenches since 2009, producing the high densities.

### 4.11.7 Recruitment estimates

There are no recruitment estimates for this FU.

### 4.11.8 MSY considerations

There is currently insufficient catch-at-length data to conduct a combined length cohort analysis, and therefore FMSY proxy harvest rates have not been calculated for this functional unit.

### 4.11.9 Short-term Forecasts

No short-term forecasts are presented for this FU.

### 4.11.10 Status of the stock

The current state of the stock is unknown.

### 4.11.11 Management considerations

The WG, ACOM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level. Management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort were compatible and in line with the scale of the resource.

There is a by-catch of other species in the Devil's Hole area. It is important that efforts are made to ensure that unwanted by-catch is kept to a minimum in this fishery. Current efforts to reduce discards and unwanted by-catches of cod under the Scottish Conservation credits scheme, include the implementation of larger meshed square mesh panels and real time closures to avoid cod.
The advice guidance and category classification for data-limited stocks (DLS) was addressed at WKLIFE2 (ICES, 2012). The methodology for DLS Nephrops stocks is further described in the 2013 Benchmark report (WKNEPH, 2013). Following the procedure outlined (section 3.1.2), an estimate of the total Nephrops grounds was used to give a likely envelope for the total abundance of Nephrops in the FU 34 (see text table below). UWTV survey information on the mean density of Nephrops ( 0.16 Nephrops $/ \mathrm{m}^{2}$ ) from the UWTV survey (2015), was used together with the mean weight (average 2007-2010) and discard percentage (average 2008-2011). The same advice as provided in 2014 of 410 tonnes (catch) results in a harvest ratio of $4.9 \%$ which is below the range of harvest ratios observed for other North Sea functional units (7.5-16\%). The 10 year average (2006-2015) results in a higher harvest ratio (8.0\%). Additional options were added to the table including an increase in $20 \%$ in catches (uncertainty cap) and a $20 \%$ reduction (precautionary buffer) on the 10 year average. These two options yield a harvest ratio of $5.9 \%$ and $6.4 \%$ respectively, both below the $7.5 \%$ threshold. The proposed advice (given in 2016) for 2017 and 2018 was that catches should be no more than 492 tonnes ( 2014 advice $+20 \%$ ). In line with the advice for other stocks, total catches, wanted catches and unwanted catches expected under the landing obligation policy were added to the table. A second catch options table allowing for discarding under the de minimis exemption ( $6 \%<$ MCRS by weight for North Sea Nephrops) is also included below. This assumes the discard patterns do not change from average values previously observed, and the total catch is split in wanted catch, unwanted catch > MCRS, unwanted catch $<$ MCRS and surviving discards. For data limited stocks the discard survival is assumed to be zero.

Catch options assuming zero discards

| Basis | Total <br> Catch | Wanted Catch | Unwanted Catch | Density (Nephrops per m2) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 0.05 | 0.1 | 0.16 | 0.2 | 0.3 | 0.4 | 0.6 | 0.8 |
| Recent average (2013-2015) | 314 | 293 | 21 | 12.08\% | 6.04\% | 3.78\% | 3.02\% | 2.01\% | 1.51\% | 1.01\% | 0.76\% |
| 2014 Advice - 20\% | 328 | 306 | 22 | 12.62\% | 6.31\% | 3.94\% | 3.16\% | 2.10\% | 1.58\% | 1.05\% | 0.79\% |
| 2014 Advice | 410 | 383 | 27 | 15.80\% | 7.90\% | 4.94\% | 3.95\% | 2.63\% | 1.97\% | 1.32\% | 0.99\% |
| 2014 Advice + $20 \%$ | 492 | 459 | 33 | 18.93\% | 9.47\% | 5.92\% | 4.73\% | 3.16\% | 2.37\% | 1.58\% | 1.18\% |
| Average (2006-2015) - $20 \%$ | 529 | 494 | 35 | 20.39\% | 10.20\% | 6.37\% | 5.10\% | 3.40\% | 2.55\% | 1.70\% | 1.27\% |
|  | 612 | 571 | 41 | 23.55\% | 11.77\% | 7.36\% | 5.89\% | 3.92\% | 2.94\% | 1.96\% | 1.47\% |
| Average (2006-2015) | 662 | 618 | 44 | 25.49\% | 12.74\% | 7.97\% | 6.37\% | 4.25\% | 3.19\% | 2.12\% | 1.59\% |
| Maximum | 1398 | 1305 | 93 | 53.82\% | 26.91\% | 16.82\% | 13.46\% | 8.97\% | 6.73\% | 4.49\% | 3.36\% |

Catch options assuming discarding allowed for de minimis excemptions

| Basis | Total Catch | Wanted Catch | Unwanted Catch >MCRS | de minimis discards <MCRS | Surviving discards | Density (Nephrops per m2) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | 0.05 | 0.1 | 0.16 | 0.2 | 0.3 | 0.4 | 0.6 | 0.8 |
| Recent average (2013-2015) | 314 | 293 | 20 | 1 | 0 | 12.08\% | 6.04\% | 3.78\% | 3.02\% | 2.01\% | 1.51\% | 1.01\% | 0.76\% |
| 2014 Advice - 20\% | 328 | 306 | 21 | 1 | 0 | 12.62\% | 6.31\% | 3.94\% | 3.16\% | 2.10\% | 1.58\% | 1.05\% | 0.79\% |
| 2014 Advice | 410 | 383 | 26 | 1 | 0 | 15.80\% | 7.90\% | 4.94\% | 3.95\% | 2.63\% | 1.97\% | 1.32\% | 0.99\% |
| 2014 Advice + 20\% | 492 | 459 | 31 | 2 | 0 | 18.93\% | 9.47\% | 5.92\% | 4.73\% | 3.16\% | 2.37\% | 1.58\% | 1.18\% |
| Average (2006-2015) - 20\% | 529 | 494 | 33 | 2 | 0 | 20.39\% | 10.20\% | 6.37\% | 5.10\% | 3.40\% | 2.55\% | 1.70\% | 1.27\% |
|  | 612 | 571 | 39 | 2 | 0 | 23.55\% | 11.77\% | 7.36\% | 5.89\% | 3.92\% | 2.94\% | 1.96\% | 1.47\% |
| Average (2006-2015) | 662 | 618 | 42 | 2 | 0 | 25.49\% | 12.74\% | 7.97\% | 6.37\% | 4.25\% | 3.19\% | 2.12\% | 1.59\% |
| Maximum | 1398 | 1305 | 88 | 5 | 0 | 53.82\% | 26.91\% | 16.82\% | 13.46\% | 8.97\% | 6.73\% | 4.49\% | 3.36\% |

Basis for the catch options.

| VARIABLE | VALUE | Source | Notes |
| :--- | :--- | :--- | :--- |
| Density in TV <br> assessment | 0.16 Nephrops <br> m 2 | ICES (2016a) | UWTV 2015 |
| Mean weight in <br> landings | 31.76 g | ICES (2016a) | Average 2007-2010 (benchmark <br> estimate WKNEPH, 2013 ) |
| Mean weight in <br> discards | 15.3 g | ICES (2016a) | Average 2013-2015 (from FU 7) |
| Mean weight in <br> unwanted catch <br> $>M C R S$ | 16.13 g | ICES (2016a) | Average 2013-2015 (from FU 7) |
| Mean weight in <br> unwanted catch <br> $<M C R S ~$ | 7.58 g | ICES (2016a) | Average 2013-2015 (from FU 7) |
| Discard rate (total) | $12.9 \%$ | ICES (2013) | Average 2008-2011 (benchmark <br> estimate WKNEPH, 2013; proportion by <br> number) |
| Discard rate <br> (>MCRS) | ICES (2016a) | Average 2013-2015 (from FU 7) |  |
| Discard rate <br> (<MCRS) | ICES (2016a) | Average 2013-2015 (from FU 7) |  |

Table 4.2.1. Nominal landings (tonnes) of Nephrops in Subarea 4, 1984-2013, as officially reported to ICES.

|  | $\begin{aligned} & \dot{\infty} \\ & \hline \\ & \hline \end{aligned}$ | $\begin{aligned} & \bullet \infty \\ & \infty \\ & \hline \end{aligned}$ | $\begin{aligned} & \circ \\ & \infty \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \hline \\ & \hline \end{aligned}$ | $\begin{aligned} & \infty \\ & \infty \\ & \hline \\ & \hline \end{aligned}$ | $\begin{aligned} & \circ \\ & 0 \\ & \hline 0 \end{aligned}$ | 옹 | $\begin{aligned} & \text { 응 } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { Nั } \\ & \text { O} \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { п교 } \\ & \hline 1 \end{aligned}$ | $\begin{aligned} & \text { オ } \\ & \hline 1 \end{aligned}$ | $\begin{aligned} & \text { ㄴN } \\ & \underset{\sim}{\circ} \end{aligned}$ | $\begin{aligned} & \circ \\ & \hline \text { ㅇ } \\ & \hline 1 \end{aligned}$ | $\begin{aligned} & \text { No} \\ & \end{aligned}$ | $\begin{aligned} & \infty \\ & \hline \mathbf{\circ} \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 638 | 679 | 344 | 437 | 500 | 574 | 610 | 427 | 384 | 418 | 304 | 410 | 185 | 311 | 238 |
| Denmark | 7 | 50 | 323 | 479 | 409 | 508 | 743 | 880 | 581 | 691 | 1128 | 1182 | 1315 | 1309 | 1440 |
| Faeroe Islands | - | - | - | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 12 | 0 | 1 | 1 |
| France | - | - | - | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Germany | . | . | . | 0 | 0 | 0 | 0 | 2 | 2 | 16 | 24 | 16 | 69 | 64 | 58 |
| Germany (Fed. Rep.) | 5 | 4 | 5 | 1 | 2 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 627 |  |
| Netherlands | - | - | - | 0 | 0 | 0 | 9 | 3 | 134 | 131 | 159 | 254 | 423 | 64 | 6945 |
| Norway | 1 | 1 | 1 | 2 | 17 | 17 | 46 | 117 | 125 | 107 | 171 | 74 | 83 | 1 | 93 |
| Sweden | - | 1 | - | 0 | 0 | 0 | 0 | 4 | 0 | 1 | 1 | 1 | 0 |  | 3 |
| UK (Eng + Wales + NI) | . | - | - | 0 | 0 | 2938 | 2332 | 1955 | 1451 | 2983 | 3613 | 2530 | 2462 | 2206 | 2094 |
| UK (Eng + Wales) | 1477 | 2052 | 2002 | 2173 | 2397 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | - | - | 8980 |
| UK (Scotland) | 4158 | 5369 | 6190 | 5304 | 6527 | 7065 | 6871 | 7501 | 6898 | 8250 | 8850 | 10018 | 8981 | 10466 | 13602 |
| UK | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |
| Total | 6286 | 8156 | 8865 | 8403 | 9852 | 11103 | 10613 | 10889 | 9575 | 12598 | 14253 | 14497 | 13518 | 15049 | 13602 |


|  | $\begin{aligned} & \text { 응 } \\ & \hline 1 \end{aligned}$ | $\begin{aligned} & \mathrm{O} \\ & \text { 아N } \end{aligned}$ | $\begin{aligned} & \bar{O} \\ & \text { N } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O} \\ & \text { N } \end{aligned}$ | $\begin{aligned} & \text { n } \\ & 0 \\ & 0 \\ & N \end{aligned}$ | $\begin{aligned} & \text { J } \\ & \text { O } \\ & \text { N } \end{aligned}$ | $\begin{aligned} & \text { n } \\ & 0 \\ & \text { N } \end{aligned}$ | $$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { N } \end{aligned}$ | $\begin{aligned} & \infty \\ & \text { O } \\ & \text { N } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { O} \\ & \text { O } \\ & \text { N } \end{aligned}$ | $$ | $\overline{\bar{O}}$ | $\sim$ $\sim$ $\sim$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 350 | 252 | 283 | 284 | 229 | 213 | 180 | 214 | 205 | 200 | 265 | 115 | 295 | 374 |
| Denmark | 1963 | 1747 | 1935 | 2154 | 2128 | 2244 | 2339 | 2024 | 1408 | 1078 | 875 | 603 | 828 | 728 |
| Faeroe Islands | 1 | 0 | - | - | - | - | - | - | - | - | - | - |  |  |
| France | 0 | 0 | - | - | - | - | - | - | - | - | - | + |  | + |
| Germany | 104 | 79 | 140 | 125 | 50 | 50 | 109 | 288 | 602 | 266 | 410 | 373 | 552 | 385 |
| Netherlands | 662 | 572 | 851 | 966 | 940 | 918 | 1019 | 982 | 1147 | 737 | 882 | 701 | 1012 | 1024 |
| Norway | 144 | 147 | 115 | 130 | 100 | 93 | 132 | 96 | 99 | 143 | 139 | 123 | 70 | 75 |
| Sweden | 4 | 37 | 26 | 14 | 1 | 1 | 3 | 1 | 5 | 26 | 2 | 1 | 1 | 1 |
| UK (Eng + Wales + NI) | 2431 | 2210 | 2691 | 1964 | 2295 | 2241 | 3236 | 4937 | 3295 | 1679 | 3437 | - |  |  |
| UK (Scotland) | 10715 | 9834 | 9681 | 11045 | 10094 | 12912 | 10565 | 16165 | 17930 | 17960 | 18587 | - |  |  |
| UK | - | - | - | - | - | - |  | - | - | - | - | 18941 | 14190 | 10976 |
| Total | 16374 | 14878 | 15722 | 16682 | 15838 | 18674 | 17583 | 24707 | 24691 | 22089 | 24597 | 20857 | 16948 | 13541 |


| Belgium | 303 | 494 | 349 |
| :---: | :---: | :---: | :---: |
| Denmark | 387 | 624 | 515 |
| Faeroe Islands | 0 | 0 | 0 |
| France | 0 | 0 | 0 |
| Germany | 425 | 418 | 435 |
| Ireland | 0 | 1 | 0 |
| Netherlands | 910 | 1154 | 1113 |
| Norway | 63 | 63 | 81 |
| Sweden | 0 |  | 0 |
| UK (Eng + Wales + NI) | - |  |  |
| UK (Scotland) | - |  |  |
| UK | 8625 | 11211 | 6825 |
| Total | 10713 | 13965 | 9318 |

* Landings data for 2015 are preliminary.

Table 3.1.2 Summary of Nephrops landings from the ICES area, by Functional Unit , 1991-2008.

| Year | FU 5 | FU 6 | FU 7 | FU 8 | FU 9 | FU 10 | FU 32 | FU 33 | $\begin{aligned} & \text { FU } \\ & 34 \end{aligned}$ | OTHER ** | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 |  | 1073 | 373 | 1006 | 1416 | 36 |  |  |  | 76 | 3980 |
| 1982 |  | 2524 | 422 | 1195 | 1120 | 19 |  |  |  | 157 | 5437 |
| 1983 |  | 2078 | 693 | 1724 | 940 | 15 |  |  |  | 101 | 5551 |
| 1984 |  | 1479 | 646 | 2134 | 1170 | 111 |  |  |  | 88 | 5628 |
| 1985 |  | 2027 | 1148 | 1969 | 2081 | 22 |  |  |  | 139 | 7386 |
| 1986 |  | 2015 | 1543 | 2263 | 2143 | 68 |  |  |  | 204 | 8236 |
| 1987 |  | 2191 | 1696 | 1674 | 1991 | 44 |  |  |  | 195 | 7791 |
| 1988 |  | 2495 | 1573 | 2528 | 1959 | 76 |  |  |  | 364 | 8995 |
| 1989 |  | 3098 | 2299 | 1886 | 2576 | 84 |  |  |  | 233 | 10176 |
| 1990 |  | 2498 | 2537 | 1930 | 2038 | 217 |  |  |  | 222 | 9442 |
| 1991 | 862 | 2063 | 4223 | 1404 | 1519 | 196 |  |  |  | 560 | 10827 |
| 1992 | 612 | 1473 | 3363 | 1757 | 1591 | 188 |  |  |  | 401 | 9385 |
| 1993 | 721 | 3030 | 3493 | 2369 | 1808 | 376 | 339 | 160 |  | 434 | 12730 |
| 1994 | 503 | 3683 | 4569 | 1850 | 1538 | 495 | 755 | 137 |  | 703 | 14233 |
| 1995 | 869 | 2569 | 6440 | 1763 | 1297 | 280 | 489 | 164 |  | 844 | 14715 |
| 1996 | 679 | 2483 | 5217 | 1688 | 1451 | 344 | 952 | 77 |  | 808 | 13699 |
| 1997 | 1149 | 2189 | 6171 | 2194 | 1446 | 316 | 760 | 276 |  | 662 | 15163 |
| 1998 | 1111 | 2177 | 5136 | 2145 | 1032 | 254 | 836 | 350 |  | 694 | 13735 |
| 1999 | 1244 | 2391 | 6521 | 2205 | 1008 | 279 | 1119 | 724 |  | 988 | 16479 |
| 2000 | 1121 | 2178 | 5569 | 1785 | 1541 | 275 | 1084 | 597 |  | 900 | 15050 |
| 2001 | 1443 | 2574 | 5541 | 1528 | 1403 | 177 | 1190 | 791 |  | 1268 | 15915 |
| 2002 | 1231 | 1954 | 7247 | 1340 | 1118 | 401 | 1170 | 861 |  | 1383 | 16705 |
| 2003 | 1144 | 2245 | 6294 | 1126 | 1079 | 337 | 1089 | 929 |  | 1390 | 15633 |
| 2004 | 1070 | 2153 | 8729 | 1658 | 1335 | 228 | 922 | 1268 |  | 1224 | 18587 |
| 2005 | 1099 | 3094 | 10685 | 1990 | 1605 | 165 | 1089 | 1050 |  | 1120 | 21897 |
| 2006 | 974 | 4903 | 10791 | 2458 | 1803 | 133 | 11033 | 1288 |  | 1249 | 24627 |
| 2007 | 1294 | 2966 | 11910 | 2652 | 1842 | 155 | 755 | 1467 |  | 1637 | 24678 |
| 2008 | 963 | 1218 | 12240 | 2450 | 1514 | 173 | 675 | 1444 |  | 1673 | 22350 |
| 2009 | 728 | 2703 | 13327 | 2662 | 1067 | 89 | 477 | 1163 |  | 2367 | 24583 |
| 2010 | 959 | 1443 | 12825 | 1871 | 1032 | 38 | 407 | 806 | 757 | $709 * * * *$ | 20847 |
| 2011 | 1053 | 2070 | 7558 | 1888 | 1391 | 69 | 395 | 1191 | 433 | $1166^{* * * * *}$ | 17214 |
| 2012 | 1240 | 2460 | 4369 | 2091 | 860 | 13 | 310 | 1084 | 597 | $608^{* * * *}$ | 13632 |
| 2013 | 1050 | 2982 | 2951 | 1503 | 623 | 16 | 191 | 946 | 120 | 409 | 10791 |
| 2014 | 1416 | 2503 | 4147 | 2370 | 1252 | 15 | 205 | 1146 | 320 | 392 | 13766 |
| 2015 | 1516 | 1371 | 1786 | 1892 | 830 | 15 | 192 | 1003 | 439 | 612 | 9656 |

* Provisional
** Includes 3.a.
***Devil's Hole landings only separated from 2011.
*** *695t in IV and 14t in 3.a
****4 only

Table 4.3.1 Nephrops in FU 5. Nominal Landings (tonnes) of Nephrops, 1991-2010, as reported to the WG.

|  | Belgium | Denmark | Netherlands | Germany | UK | Total** | CATCH*** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 682 | 176 | na |  | 4 | 862 |  |
| 1992 | 571 | 22 | na |  | 19 | 612 |  |
| 1993 | 694 | 20 | na |  | 7 | 721 |  |
| 1994 | 494 | 0 | na |  | 9 | 503 |  |
| 1995 | 641 | 77 | 148 |  | 3 | 869 |  |
| 1996 | 266 | 41 | 317 |  | 55 | 679 |  |
| 1997 | 486 | 67 | 540 |  | 56 | 1149 |  |
| 1998 | 372 | 88 | 584 | 39 | 28 | 1111 |  |
| 1999 | 436 | 53 | 538 | 59 | 158 | 1244 |  |
| 2000 | 366 | 83 | 402 | 52 | 218 | 1121 |  |
| 2001 | 353 | 145 | 553 | 114 | 278 | 1443 |  |
| 2002 | 281 | 94 | 617 | 88 | 151 | 1231 |  |
| 2003 | 265 | 36 | 661 | 24 | 158 | 1144 |  |
| 2004 | 171 | 39 | 646 | 16 | 198 | 1070 |  |
| 2005 | 109 | 87 | 654 | 51 | 198 | 1099 |  |
| 2006 | 77 | 24 | 444 | 99 | 330 | 974 |  |
| 2007 | 75 | 3 | 464 | 201 | 551 | 1294 |  |
| 2008 | 49 | 29 | 268 | 108 | 509 | 963 |  |
| 2009 | 52 | 3 | 288 | 98 | 287 | 728 |  |
| 2010 | 48 | 5 | 354 | 140 | 411 | 959 |  |
| 2011 | 60 | 18 | 480 | 145 | 350 | 1053 |  |
| 2012 | 129 | 0 | 497 | 121 | 493 | 1240 |  |
| 2013 | 142 | 1 | 447 | 168 | 292 | 1050 |  |
| 2014 | 131 | 41 | 645 | 139 | 460 | 1416 |  |
| 2015 | 146 | 0 | 681 | 184 | 505 | 1516 | 3562 |

[^1]Table 4.3.2. Nephrops in FU5. Mean length (mm) by sex in landings from Dutch sampling.

|  |  | Mean LenGth (мM) |
| :--- | :--- | :--- |
| Year | Females | Males |
| 2003 | 38.43 | 38.43 |
| 2004 | 37.68 | 39.21 |
| 2005 | 36.85 | 37.47 |
| 2006 | 37.33 | 37.85 |
| 2007 | 38.05 | 38.9 |
| 2008 | 38.71 | 39.81 |
| 2009 | 38.18 | 39.91 |
| 2010 | 41.1 | 41.1 |
| 2011 | 41.2 | 41.1 |
| 2012 | 39.7 | 40.8 |
| 2013 | na | na |
| 2014 | 40.2 | 40.2 |
| 2015 | 39.43 | 39.8 |

Table 4.3.3 Nephrops in FU5. Landings, effort and LPUE for directed fisheries.

| ENGLAND |  |  |  |
| :--- | :--- | :--- | :--- |
|  | Landings | Effort | LPUE |
|  | tons | Boat Days Fished | T/Day |
| 2000 | 53.2184 | 36 | 1.48 |
| 2001 | 104.1648 | 73 | 1.43 |
| 2002 | 7.3549 | 10 | 0.74 |
| 2003 | 21.4591 | 24 | 0.89 |
| 2004 | 32.4969 | 21 | 1.55 |
| 2005 | 66.7731 | 35 | 1.91 |
| 2006 | 176.7924 | 214 | 0.83 |
| 2007 | 208.698 | 177 | 1.18 |
| 2008 | 267.7608 | 292 | 0.92 |
| 2009 | 193.9114 | 188 | 1.03 |
| 2010 | 176.1818 | 152 | 1.16 |
| 2011 | 181.6175 | 147 | 1.24 |
| 2012 | 204.7108 | 185 | 1.11 |
| 2013 | 111.6035 | 142 | 0.79 |
| 2014 | 147.0582 | 138 | 1.07 |
| 2015 | 136.1702 | 147 | 0.93 |

* provisional na = not available

Logbook records from vessels operating in FU 5 , with mesh size $>=70 \mathrm{~mm}$ with Nephrops in catches

Table 4.4.1 Nephrops in FU 6. Nominal Landings (tonnes) of Nephrops, 1981-2015, as reported to the WG.

| Year | UK England \& N. IreLAND | UK Scotland | Sub total | Other COUNTRIES** | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 1006 | 67 | 1073 | 0 | 1073 |
| 1982 | 2443 | 81 | 2524 | 0 | 2524 |
| 1983 | 2073 | 5 | 2078 | 0 | 2078 |
| 1984 | 1471 | 8 | 1479 | 0 | 1479 |
| 1985 | 2009 | 18 | 2027 | 0 | 2027 |
| 1986 | 1987 | 28 | 2015 | 0 | 2015 |
| 1987 | 2158 | 33 | 2191 | 0 | 2191 |
| 1988 | 2390 | 105 | 2495 | 0 | 2495 |
| 1989 | 2930 | 168 | 3098 | 0 | 3098 |
| 1990 | 2306 | 192 | 2498 | 0 | 2498 |
| 1991 | 1884 | 179 | 2063 | 0 | 2063 |
| 1992 | 1403 | 60 | 1463 | 10 | 1473 |
| 1993 | 2941 | 89 | 3030 | 0 | 3030 |
| 1994 | 3530 | 153 | 3683 | 0 | 3683 |
| 1995 | 2478 | 90 | 2568 | 1 | 2569 |
| 1996 | 2386 | 96 | 2482 | 1 | 2483 |
| 1997 | 2109 | 80 | 2189 | 0 | 2189 |
| 1998 | 2029 | 147 | 2176 | 1 | 2177 |
| 1999 | 2197 | 194 | 2391 | 0 | 2391 |
| 2000 | 1947 | 231 | 2178 | 0 | 2178 |
| 2001 | 2319 | 255 | 2574 | 0 | 2574 |
| 2002 | 1739 | 215 | 1954 | 0 | 1954 |
| 2003 | 2031 | 214 | 2245 | 0 | 2245 |
| 2004 | 1952 | 201 | 2153 | 0 | 2153 |
| 2005 | 2936 | 158 | 3094 | 0 | 3094 |
| 2006 | 4430 | 434 | 4864 | 39 | 4903 |
| 2007 | 2525 | 437 | 2962 | 4 | 2966 |
| 2008 | 976 | 244 | 1220 | 0 | 1220 |
| 2009 | 2299 | 414 | 2713 | 0 | 2713 |
| 2010 | 1258 | 185 | 1443 | 0 | 1443 |
| 2011 | 1806 | 250 | 2056 | 14 | 2070 |
| 2012 | 2177 | 256 | 2433 | 27 | 2460 |
| 2013 | 2666 | 305 | 2971 | 11 | 2982 |
| 2014 | 2104 | 345 | 2449 | 54 | 2503 |
| 2015* | 1186 | 174 | 1360 | 11 | 1371 |

* provisional na = not available
** Other countries includes Ne, Be and Dk

Table 4.4.2: Nephrops in FU 6: Landings and effort by English vessels targeting Nephrops.

| Year | $<10 \mathrm{M}$ |  |  | 10-15m |  |  | >15M |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Landings | Days | LPUE <br> (kg/d) | Landings | Days | LPUE <br> (kg/d) | Landings | Days | LPUE <br> (kg/d) |
| 2000 | 124 | 591 | 210 | 368 | 1611 | 228 | 552 | 1465 | 377 |
| 2001 | 139 | 665 | 209 | 306 | 1264 | 242 | 460 | 1363 | 338 |
| 2002 | 125 | 654 | 191 | 354 | 1376 | 257 | 456 | 1320 | 346 |
| 2003 | 319 | 958 | 333 | 483 | 1614 | 299 | 517 | 1461 | 354 |
| 2004 | 384 | 1088 | 353 | 456 | 1604 | 284 | 371 | 863 | 430 |
| 2005 | 581 | 1472 | 395 | 511 | 1669 | 306 | 647 | 1276 | 507 |
| 2006 | 778 | 2296 | 339 | 489 | 1372 | 356 | 1324 | 2062 | 642 |
| 2007 | 523 | 2067 | 253 | 259 | 1034 | 251 | 568 | 1571 | 362 |
| 2008 | 299 | 2181 | 137 | 152 | 798 | 190 | 163 | 611 | 266 |
| 2009 | 449 | 2279 | 197 | 314 | 1103 | 285 | 574 | 1195 | 480 |
| 2010 | 340 | 1773 | 192 | 176 | 650 | 271 | 322 | 969 | 332 |
| 2011 | 401 | 2320 | 173 | 235 | 827 | 285 | 414 | 1006 | 412 |
| 2012 | 388 | 2174 | 178 | 333 | 1263 | 264 | 406 | 1014 | 400 |
| 2013 | 465 | 2374 | 196 | 402 | 1246 | 323 | 484 | 899 | 539 |
| 2014 | 399 | 2160 | 185 | 280 | 870 | 322 | 420 | 917 | 458 |
| 2015 | 197 | 1567 | 125 | 126 | 647 | 195 | 242 | 901 | 269 |

Table 4.4.3 Nephrops in FU 6: Mean sizes in catches and landings by sex.

| Year | Catches |  | LaNDINGS |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Males | Females | Males | Females |
| 1985 | 30.1 | 28.5 | 35.4 | 33.8 |
| 1986 | 31.7 | 30.2 | 35.3 | 33.7 |
| 1987 | 28.6 | 27 | 35.3 | 33.3 |
| 1988 | 28.7 | 27.3 | 35 | 33.9 |
| 1989 | 29 | 28.2 | 32.4 | 31.9 |
| 1990 | 27.1 | 27.4 | 31.8 | 31.3 |
| 1991 | 28.9 | 27.1 | 33.5 | 33.1 |
| 1992 | 30.8 | 29 | 33 | 31.9 |
| 1993 | 32.1 | 28.7 | 33.4 | 30.1 |
| 1994 | 30.5 | 27.7 | 33.8 | 30.5 |
| 1995 | 28.4 | 27.4 | 33.8 | 31.6 |
| 1996 | 29.8 | 28.2 | 34.5 | 32.1 |
| 1997 | 29.9 | 29.6 | 33.5 | 32.1 |
| 1998 | 30 | 28.9 | 34.9 | 33.7 |
| 1999 | 29.6 | 27.5 | 35.1 | 33.6 |
| 2000 | 27.2 | 26.8 | 31.1 | 31.3 |
| 2001 | 26.2 | 26.3 | 30.6 | 31.3 |
| 2002 | 28.0 | 26.9 | 30.9 | 30.0 |
| 2003 | 29.0 | 27.1 | 31.7 | 30.6 |
| 2004 | 29.2 | 27.0 | 32.3 | 30.6 |
| 2005 | 29.7 | 29.4 | 32.1 | 32.2 |
| 2006 | 29.0 | 30.3 | 31.4 | 32.4 |
| 2007 | 31.3 | 30.7 | 33.3 | 32.6 |
| 2008 | 31.5 | 31.1 | 33.5 | 33.3 |
| 2009 | 30.0 | 31.0 | 32.1 | 33.3 |
| 2010 | 31.2 | 31.4 | 32.8 | 33.2 |
| 2011 | 32.0 | 31.6 | 33.7 | 33.6 |
| 2012 | 30.8 | 32.0 | 33.2 | 34.5 |
| 2013 | 29.6 | 32.4 | 32.0 | 35.3 |
| 2014 | 31.8 | 35.4 | 32.9 | 36.6 |
| 2015 | 31.5 | 31.7 | 33.9 | 34.9 |
| * provisional na = not available |  |  |  |  |

Table 4.4.4 Nephrops in FU 6: Results of the UWTV survey.

| Year | Stations | SEASON | Mean density | Absolute <br> Abundance | 95\% <br> CONFIDENCE INTERVAL | Method |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | burrows/m ${ }^{2}$ | millions | millions |  |
| 1997 | 87 | Autumn | 0.46 | 1500 | 125 | Box |
| 1998 | 91 | Autumn | 0.33 | 1090 | 89 | Box |
| 1999 | - | Autumn | No survey |  |  | Box |
| 2000 | - | Autumn | No survey |  |  | Box |
| 2001 | 180 | Autumn | 0.56 | 1685 | 67 | Box |
| 2002 | 37 | Autumn | 0.33 | 1048 | 112 | Box |
| 2003 | 73 | Autumn | 0.33 | 1085 | 90 | Box |
| 2004 | 76 | Autumn | 0.43 | 1377 | 101 | Box |
| 2005 | 105 | Autumn | 0.49 | 1657 | 148 | Box |
| 2006 | 105 | Autumn* | 0.37 | 1244 | 114 | Box |
| 2007 | 105 | Autumn* | 0.28 | 858 | 23 | Geostatistics |
| 2008 | 95 | Autumn* | 0.31 | 987 | 39 | Geostatistics |
| 2009 | 76 | Autumn* | 0.22 | 682 | 38 | Geostatistics |
| 2010 | 95 | Autumn* | 0.25 | 785 | 21 | Geostatistics |
| 2011 | 97 | Autumn* | 0.28 | 878 | 17 | Geostatistics |
| 2012 | 97 | Autumn* | 0.24 | 758 | 13 | Geostatistics |
| 2013 | 110 | Summer | 0.23 | 706 | 18 | Geostatistics |
| 2014 | 110 | Summer | 0.24 | 755 | 18 | Geostatistics |
| 2015 | 110 | Summer | 0.18 | 565 | 13 | Geostatistics |

Table 4.4.5 Nephrops in FU 6: Historical harvest rate determination.


Table 4.5.1 Nephrops, Fladen (FU 7), Nominal Landings (tonnes) of Nephrops, 1981-2015, as reported to the WG.

|  | UK Scotland |  |  |  |  | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Nephrops | Other | Creel | Sub-total | Denmark | countries |  |
| Year | trawl | trawl |  |  |  | ** |  |
| 1981 | 304 | 68 | 0 | 372 | 0 | 0 | 372 |
| 1982 | 381 | 40 | 0 | 421 | 0 | 0 | 421 |
| 1983 | 588 | 105 | 0 | 693 | 0 | 0 | 693 |
| 1984 | 552 | 94 | 0 | 646 | 0 | 0 | 646 |
| 1985 | 1020 | 120 | 0 | 1140 | 7 | 0 | 1147 |
| 1986 | 1401 | 92 | 0 | 1493 | 50 | 0 | 1543 |
| 1987 | 1023 | 349 | 0 | 1372 | 323 | 0 | 1695 |
| 1988 | 1309 | 185 | 0 | 1494 | 81 | 0 | 1575 |
| 1989 | 1724 | 410 | 0 | 2134 | 165 | 0 | 2299 |
| 1990 | 1703 | 598 | 0 | 2301 | 236 | 3 | 2540 |
| 1991 | 3021 | 772 | 0 | 3793 | 424 | 6 | 4223 |
| 1992 | 1809 | 1164 | 0 | 2973 | 359 | 31 | 3363 |
| 1993 | 2031 | 1234 | 0 | 3265 | 224 | 3 | 3492 |
| 1994 | 1816 | 2356 | 0 | 4172 | 390 | 6 | 4568 |
| 1995 | 3568 | 2389 | 19 | 5976 | 439 | 4 | 6419 |
| 1996 | 2338 | 2578 | 7 | 4923 | 286 | 1 | 5210 |
| 1997 | 2712 | 3221 | 0 | 5933 | 235 | 2 | 6170 |
| 1998 | 2290 | 2673 | 0 | 4963 | 173 | 0 | 5136 |
| 1999 | 2860 | 3546 | 0 | 6406 | 96 | 16 | 6518 |
| 2000 | 2916 | 2546 | 0 | 5462 | 103 | 5 | 5570 |
| 2001 | 3540 | 1936 | 0 | 5476 | 64 | 2 | 5542 |
| 2002 | 4511 | 2546 | 0 | 7057 | 173 | 15 | 7245 |
| 2003 | 4175 | 2033 | 0 | 6208 | 82 | 4 | 6294 |
| 2004 | 7274 | 1319 | 1 | 8594 | 136 | 0 | 8730 |
| 2005 | 8849 | 1508 | 5 | 10362 | 321 | 1 | 10684 |
| 2006 | 9470 | 1026 | 1 | 10497 | 283 | 11 | 10791 |
| 2007 | 11055 | 734 | 0 | 11789 | 119 | 3 | 11911 |
| 2008 | 11432 | 666 | 0 | 12098 | 133 | 8 | 12239 |
| 2009 | 12688 | 499 | 0 | 13187 | 130 | 10 | 13327 |
| 2010 | 12544 | 288 | 0 | 12832 | 124 | 12 | 12968 |
| 2011 | 7367 | 128 | 0 | 7495 | 64 | <0.5 | 7559 |
| 2012 | 4257 | 81 | 0 | 4338 | 75 | 2 | 4415 |
| 2013 | 2275 | 663 | 0 | 2938 | 5 | 8 | 2951 |
| 2014 | 2164 | 1970 | 0 | 4134 | 10 | 3 | 4147 |
| 2015 | 806 | 968 | 0 | 1774 | 8 | 4 | 1786 |
| * provisional na = not available |  |  |  |  |  |  |  |
| **Other countries includes Belgium, Norway, Sweden and UK England |  |  |  |  |  |  |  |

Table 4.5.2 Nephrops, Fladen (FU 7): landings, effort (days fishing) and LPUE (kg/day) for UK bottom trawlers landing in Scotland and fishing Nephrops with codend mesh sizes of 70 mm or above, 2000-2015.

|  | Year | LANDINGS (TONNES) | Effort (DAYS) |
| :--- | :--- | :--- | :--- |
| 2000 | 5462 | 35367 | LPUE (KG/DAY) |
| 2001 | 5476 | 28558 | 154.4 |
| 2002 | 7057 | 28586 | 191.8 |
| 2003 | 6208 | 21960 | 246.9 |
| 2004 | 8593 | 21562 | 282.7 |
| 2005 | 10357 | 23555 | 398.5 |
| 2006 | 10496 | 22836 | 439.7 |
| 2007 | 11789 | 21603 | 459.6 |
| 2008 | 12098 | 22856 | 545.7 |
| 2009 | 13187 | 21153 | 529.3 |
| 2010 | 12832 | 20968 | 623.4 |
| 2011 | 7495 | 15273 | 612.0 |
| 2012 | 4338 | 11994 | 490.7 |
| 2013 | 2938 | 11933 | 361.7 |
| 2014 | 4134 | 12629 | 246.2 |
| $2015^{*}$ | 1774 | 10562 | 327.3 |

* provisional

Table 4.5.3 Nephrops, Fladen (FU 7): Logbook recorded effort (kW days) and LPUE (kg/kW day) for bottom trawlers catching Nephrops with codend mesh sizes of 70 mm or above, and estimated total effort by Danish trawlers, 1991-2015.

| Year | LOGBOOK DATA |  |
| :---: | :---: | :---: |
|  | Effort | LPUE |
| 1991 | 2487464 | 0.170 |
| 1992 | 1952431 | 0.184 |
| 1993 | 653665 | 0.343 |
| 1994 | 1029253 | 0.379 |
| 1995 | 696951 | 0.630 |
| 1996 | 524375 | 0.545 |
| 1997 | 278210 | 0.845 |
| 1998 | 207196 | 0.835 |
| 1999 | 144720 | 0.663 |
| 2000 | 236941 | 0.435 |
| 2001 | 142562 | 0.449 |
| 2002 | 217053 | 0.797 |
| 2003 | 105864 | 0.775 |
| 2004 | 196984 | 0.690 |
| 2005 | 430272 | 0.746 |
| 2006 | 363866 | 0.778 |
| 2007 | 160590 | 0.741 |
| 2008 | 106969 | 1.243 |
| 2009 | 92461 | 1.406 |
| 2010 | 125830 | 0.985 |
| 2011 | 65646 | 0.975 |
| 2012 | 129719 | 0.578 |
| 2013 | 130458 | 0.038 |
| 2014 | 171105 | 0.058 |
| 2015 | 71790 | 0.111 |

Table 4.5.4 Nephrops, Fladen (FU 7): Mean sizes (CL mm) above and below 35 mm of male and female Nephrops in Scottish catches and landings, 1993-2015.

| Year | Catches |  | LANDINGS |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | < 35 mm CL |  | < 35 mm CL |  | $>35 \mathrm{~mm} \mathrm{CL}$ |  |
|  | Males | Females | Males | Females | Males | Females |
| 1993 | na | na | 30.4 | 29.6 | 38.7 | 38.2 |
| 1994 | na | na | 30.0 | 28.9 | 39.2 | 37.8 |
| 1995 | na | na | 30.6 | 29.8 | 39.9 | 38.1 |
| 1996 | na | na | 30.4 | 29.1 | 40.6 | 38.8 |
| 1997 | na | na | 30.2 | 29.1 | 40.9 | 38.8 |
| 1998 | na | na | 30.8 | 29.4 | 40.7 | 38.3 |
| 1999 | na | na | 30.9 | 29.6 | 40.5 | 38.5 |
| 2000 | 30.7 | 30.1 | 31.2 | 30.5 | 41.3 | 38.7 |
| 2001 | 30.1 | 29.4 | 30.7 | 29.7 | 39.6 | 38.0 |
| 2002 | 30.6 | 30.0 | 31.3 | 30.7 | 39.5 | 38.3 |
| 2003 | 30.9 | 29.8 | 31.2 | 30.1 | 40.0 | 38.1 |
| 2004 | 30.8 | 29.9 | 31.1 | 30.2 | 40.1 | 38.7 |
| 2005 | 30.9 | 30.0 | 31.2 | 30.1 | 40.1 | 38.2 |
| 2006 | 30.3 | 29.7 | 30.8 | 30.0 | 40.7 | 38.2 |
| 2007 | 29.8 | 29.2 | 30.4 | 29.5 | 40.8 | 38.8 |
| 2008 | 29.7 | 28.6 | 29.8 | 28.7 | 41.8 | 39.1 |
| 2009 | 30.7 | 29.5 | 31.2 | 29.9 | 39.7 | 38.7 |
| 2010 | 30.4 | 29.0 | 30.5 | 29.0 | 39.8 | 38.4 |
| 2011 | 31.7 | 29.6 | 31.7 | 29.6 | 41.2 | 38.6 |
| 2012 | 31.9 | 30.6 | 31.9 | 30.6 | 41.8 | 38.5 |
| 2013 | 31.4 | 30.2 | 31.4 | 30.2 | 42.2 | 39.0 |
| 2014 | 30.4 | 30.1 | 30.8 | 30.2 | 44.5 | 39.2 |
| 2015 | 32.3 | 31.2 | 32.3 | 31.2 | 41.5 | 40.0 |
| na = not available |  |  |  |  |  |  |

Table 4.5.5 Nephrops, FUs 7-9 and 34 (Fladen, Firth of Forth, Moray Firth and Devil's Hole. Mean weight ( g ) in the landings.

| Year | Fladen | FIRTH of Forth | Moray Firth | Devil's Hole | Noup |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 31.59 | 20.29 | 20.05 | na | na |
| 1991 | 26.50 | 20.03 | 18.53 | na | na |
| 1992 | 29.61 | 20.96 | 23.49 | na | na |
| 1993 | 25.38 | 24.30 | 23.42 | na | na |
| 1994 | 23.72 | 19.51 | 22.25 | na | na |
| 1995 | 27.51 | 19.55 | 20.59 | na | na |
| 1996 | 29.82 | 20.81 | 21.40 | na | na |
| 1997 | 32.08 | 18.87 | 20.43 | na | 23.94 |
| 1998 | 31.37 | 18.23 | 20.47 | na | 20.58 |
| 1999 | 30.55 | 20.05 | 21.79 | na | 21.23 |
| 2000 | 36.35 | 21.83 | 25.44 | na | 30.81 |
| 2001 | 25.10 | 21.22 | 24.18 | na | 25.30 |
| 2002 | 27.93 | 19.62 | 27.68 | na | 27.95 |
| 2003 | 30.15 | 22.31 | 23.32 | na | 20.05 |
| 2004 | 30.98 | 22.45 | 27.57 | na | 28.98 |
| 2005 | 29.05 | 22.33 | 23.84 | na | 24.13 |
| 2006 | 29.25 | 21.43 | 22.34 | 22.93 | 25.97 |
| 2007 | 26.63 | 20.97 | 23.04 | 26.27 | 25.58 |
| 2008 | 28.18 | 17.23 | 25.29 | 30.08 | 33.18 |
| 2009 | 28.20 | 19.41 | 23.46 | 39.62 | 49.38 |
| 2010 | 26.38 | 19.76 | 26.94 | 31.08 | 51.93 |
| 2011 | 36.17 | 19.75 | 21.63 | 42.05 | 45.73 |
| 2012 | 36.91 | 21.66 | 23.16 | na | 34.48 |
| 2013 | 34.90 | 19.30 | 24.95 | na | 43.56 |
| 2014 | 43.11 | 24.30 | 28.94 | 50.09 | 68.31 |
| 2015 | 36.70 | 21.84 | 29.10 | 48.75 | na |
| Mean (13-15) | 38.24 | 21.81 | 27.66 | 31.76* | - |

* Mean weight for Devil's Hole based on 2007-2010 range (WKNEPH, 2013)
na $=$ not available

Table 4.5.6. Nephrops, Fladen (FU 7): Results of the 1992-2015 TV surveys.
$\left.\begin{array}{lllll}\hline & & & & \text { MEAN DENSITY }\end{array} \quad \begin{array}{c}\text { 95\% CONFIDENCE } \\ \text { INTERVAL }\end{array}\right)$

Table 4.5.7. Nephrops, Fladen Ground (FU 7):Summary of TV results for most recent 3 years (20132015) showing strata surveyed, numbers of stations in each strata, mean density and observed variance, overall abundance and variance raised to stratum area. Proportion indicates relative amounts of overall raised variance attributable to each stratum.

| Stratum (RANGES OF \% SILT CLAY) | Area (км2) | Number <br> OF <br> Stations | Mean <br> BURROW DENSITY (NO./M2) | Observed VARIANCE | Abundance (MILLIONS) | Stratum Variance | Proportion of total VARIANCE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 TV survey |  |  |  |  |  |  |  |
| $>80$ | 3248 | 9 | 0.183 | 0.001 | 593 | 1867 | 0.049 |
| 55<80 | 4967 | 14 | 0.166 | 0.007 | 823 | 11739 | 0.309 |
| 40<55 | 4304 | 15 | 0.115 | 0.005 | 493 | 5982 | 0.158 |
| <40 | 15634 | 33 | 0.064 | 0.002 | 993 | 18347 | 0.484 |
| Total | 28153 | 71 |  |  | 2902 | 37934 | 1 |
| 2014 TV survey |  |  |  |  |  |  |  |
| $>80$ | 3248 | 9 | 0.197 | 0.007 | 639 | 7993 | 0.188 |
| $55<80$ | 4967 | 15 | 0.143 | 0.004 | 709 | 6387 | 0.15 |
| 40<55 | 4304 | 12 | 0.112 | 0.004 | 481 | 6643 | 0.156 |
| <40 | 15634 | 34 | 0.074 | 0.003 | 1162 | 21432 | 0.505 |
| Total | 28153 | 70 |  |  | 2990 | 42455 | 1 |
| 2015 TV survey |  |  |  |  |  |  |  |
| $>80$ | 3248 | 10 | 0.201 | 0.002 | 652 | 2450 | 0.096 |
| $55<80$ | 4967 | 15 | 0.124 | 0.002 | 613 | 4043 | 0.158 |
| $40<55$ | 4304 | 12 | 0.096 | 0.004 | 414 | 6174 | 0.241 |
| <40 | 15634 | 34 | 0.057 | 0.002 | 889 | 12974 | 0.506 |
| Total | 28153 | 71 |  |  | 2569 | 25642 | 1 |

Table 4.5.8 Nephrops, Fladen (FU 7): Adjusted TV survey abundance, landings, total discard rate (proportion by number), dead discard rate and estimated harvest ratio 2003-2015.


Table 4.6.1 Nephrops, Firth of Forth (FU 8), Nominal Landings (tonnes) of Nephrops, 1981-2015, as reported to the WG.

| Year | UK Scotland |  |  |  | UK |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Other |  |  |  <br> NI) |  |
|  | Nephrops trawl | trawl | Creel | Sub-total |  | Total ** |
| 1981 | 947 | 60 | 0 | 1007 | 0 | 1007 |
| 1982 | 1138 | 57 | 0 | 1195 | 0 | 1195 |
| 1983 | 1681 | 43 | 0 | 1724 | 0 | 1724 |
| 1984 | 2078 | 56 | 0 | 2134 | 0 | 2134 |
| 1985 | 1907 | 61 | 0 | 1968 | 0 | 1968 |
| 1986 | 2204 | 59 | 0 | 2263 | 0 | 2263 |
| 1987 | 1583 | 90 | 2 | 1675 | 0 | 1675 |
| 1988 | 2455 | 74 | 0 | 2529 | 0 | 2529 |
| 1989 | 1834 | 53 | 0 | 1887 | 1 | 1888 |
| 1990 | 1900 | 30 | 0 | 1930 | 1 | 1931 |
| 1991 | 1362 | 43 | 0 | 1405 | 0 | 1405 |
| 1992 | 1715 | 41 | 0 | 1756 | 0 | 1756 |
| 1993 | 2349 | 17 | 0 | 2366 | 2 | 2368 |
| 1994 | 1827 | 17 | 0 | 1844 | 6 | 1850 |
| 1995 | 1707 | 53 | 0 | 1760 | 2 | 1762 |
| 1996 | 1621 | 66 | 0 | 1687 | 0 | 1687 |
| 1997 | 2136 | 55 | 0 | 2191 | 2 | 2193 |
| 1998 | 2105 | 37 | 0 | 2142 | 2 | 2144 |
| 1999 | 2193 | 10 | 1 | 2204 | 3 | 2207 |
| 2000 | 1775 | 9 | 0 | 1784 | 1 | 1785 |
| 2001 | 1484 | 34 | 0 | 1518 | 9 | 1527 |
| 2002 | 1302 | 31 | 1 | 1334 | 6 | 1340 |
| 2003 | 1116 | 8 | 0 | 1124 | 3 | 1127 |
| 2004 | 1650 | 4 | 0 | 1654 | 3 | 1657 |
| 2005 | 1974 | 0 | 4 | 1978 | 11 | 1989 |
| 2006 | 2438 | 3 | 12 | 2453 | 5 | 2458 |
| 2007 | 2627 | 10 | 7 | 2644 | 7 | 2651 |
| 2008 | 2435 | 2 | 8 | 2445 | 5 | 2450 |
| 2009 | 2620 | 8 | 26 | 2654 | 9 | 2663 |
| 2010 | 1923 | 5 | 13 | 1941 | 9 | 1950 |
| 2011 | 1789 | 6 | 89 | 1884 | 5 | 1889 |
| 2012 | 1944 | 17 | 126 | 2087 | 42 | 2129 |
| 2013 | 1409 | 24 | 58 | 1491 | 12 | 1503 |
| 2014 | 2313 | 33 | 14 | 2360 | 22 | 2382 |
| 2015* | 1677 | 104 | 43 | 1824 | 68 | 1892 |

[^2]Table 4.6.2 Nephrops, Firth of Forth (FU 8): landings, effort (days fishing) and LPUE (kg/day) for UK bottom trawlers landing in Scotland and fishing Nephrops with codend mesh sizes of 70 mm or above, 2000-2015

|  | Year | LANDINGS (tonNes) | Effort (DAYS) |
| :--- | :--- | :--- | :--- |
| 2000 | 1784 | 10508 | LPUE (KG/DAY) |
| 2001 | 1518 | 11513 | 169.8 |
| 2002 | 1333 | 10394 | 131.9 |
| 2003 | 1124 | 8279 | 128.2 |
| 2004 | 1654 | 9505 | 135.8 |
| 2005 | 1974 | 7704 | 174.0 |
| 2006 | 2441 | 6174 | 256.2 |
| 2007 | 2637 | 6409 | 395.4 |
| 2008 | 2437 | 6440 | 411.5 |
| 2009 | 2628 | 5852 | 378.4 |
| 2010 | 1928 | 5054 | 449.1 |
| 2011 | 1795 | 4614 | 381.5 |
| 2012 | 1961 | 5058 | 389.0 |
| 2013 | 1433 | 4029 | 387.7 |
| 2014 | 2346 | 6812 | 355.7 |
| $2015^{*}$ | 1781 | 6024 | 344.4 |

* provisional na = not available

Table 4.6.3 Nephrops, Firth of Forth (FU 8): Mean sizes (CL mm) above and below 35 mm of male and female Nephrops in Scottish catches and landings, 1981-2015

| Year | Catches |  | Landings |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | < 35 mm CL |  | < 35 mm CL |  | $>35 \mathrm{~mm} \mathrm{CL}$ |  |
|  | Males | Females | Males | Females | Males | Females |
| 1981 | na | na | 31.5 | 31.0 | 39.7 | 38.7 |
| 1982 | na | na | 30.4 | 30.1 | 40.0 | 39.1 |
| 1983 | na | na | 31.1 | 30.8 | 40.2 | 38.7 |
| 1984 | na | na | 30.3 | 29.7 | 39.4 | 38.4 |
| 1985 | na | na | 30.6 | 29.9 | 39.4 | 38.2 |
| 1986 | na | na | 29.7 | 29.2 | 39.1 | 38.5 |
| 1987 | na | na | 29.9 | 29.6 | 39.1 | 38.2 |
| 1988 | na | na | 28.5 | 28.5 | 39.1 | 39.0 |
| 1989 | na | na | 29.2 | 28.9 | 38.7 | 38.9 |
| 1990 | 28.9 | 27.8 | 29.8 | 28.6 | 38.3 | 38.8 |
| 1991 | 28.7 | 27.5 | 29.8 | 28.7 | 38.3 | 38.7 |
| 1992 | 29.5 | 27.9 | 30.2 | 28.7 | 38.1 | 38.7 |
| 1993 | 28.7 | 28.0 | 30.3 | 29.5 | 39.0 | 38.6 |
| 1994 | 25.7 | 25.1 | 29.1 | 28.5 | 38.8 | 37.8 |
| 1995 | 27.9 | 27.1 | 29.4 | 28.9 | 38.7 | 37.9 |
| 1996 | 28.0 | 27.4 | 29.8 | 28.8 | 38.6 | 38.6 |
| 1997 | 27.2 | 27.0 | 29.2 | 28.7 | 38.8 | 38.2 |
| 1998 | 27.7 | 26.4 | 29.0 | 27.9 | 38.5 | 38.4 |
| 1999 | 27.2 | 26.5 | 29.6 | 28.8 | 38.0 | 37.9 |
| 2000 | 28.5 | 27.2 | 30.6 | 29.8 | 38.2 | 38.3 |
| 2001 | 28.1 | 27.0 | 30.6 | 29.2 | 38.0 | 37.9 |
| 2002 | 27.1 | 26.3 | 29.8 | 29.3 | 38.3 | 37.9 |
| 2003 | 27.2 | 25.4 | 30.2 | 29.1 | 38.1 | 38.0 |
| 2004 | 28.6 | 27.8 | 30.7 | 30.0 | 38.4 | 37.6 |
| 2005 | 27.6 | 26.9 | 30.3 | 30.0 | 38.7 | 38.2 |
| 2006 | 27.3 | 27.0 | 29.8 | 29.9 | 38.7 | 37.8 |
| 2007 | 29.2 | 28.3 | 29.8 | 28.6 | 39.1 | 38.6 |
| 2008 | 27.7 | 27.2 | 28.1 | 26.9 | 39.4 | 37.9 |
| 2009 | 27.5 | 26.2 | 29.7 | 28.5 | 38.3 | 38.0 |
| 2010 | 28.3 | 26.9 | 29.8 | 28.4 | 38.6 | 38.2 |
| 2011 | 28.6 | 27.5 | 30.0 | 28.3 | 38.8 | 38.2 |
| 2012 | 28.4 | 28.0 | 30.4 | 29.3 | 39.0 | 38.1 |
| 2013 | 28.3 | 27.4 | 29.6 | 28.8 | 38.8 | 37.9 |
| 2014 | 29.6 | 29.1 | 31.1 | 30.3 | 38.6 | 38.1 |
| 2015 | 27.9 | 28.3 | 29.5 | 29.3 | 39.6 | 38.5 |
| na = not available |  |  |  |  |  |  |

Table 4.6.4. Nephrops, Firth of Forth (FU 8): Results of the 1993-2015 TV surveys.

| Year | Stations | Mean Density | Abundance | 95\% CONF INTERVAL |
| :---: | :---: | :---: | :---: | :---: |
|  |  | burrows/m ${ }^{2}$ | millions | millions |
| 1993 | 37 | 0.61 | 555 | 142 |
| 1994 | 30 | 0.49 | 448 | 78 |
| 1995 | no survey |  |  |  |
| 1996 | 27 | 0.41 | 375 | 88 |
| 1997 | no survey |  |  |  |
| 1998 | 32 | 0.32 | 292 | 81 |
| 1999 | 49 | 0.51 | 463 | 78 |
| 2000 | 53 | 0.48 | 443 | 70 |
| 2001 | 46 | 0.46 | 419 | 79 |
| 2002 | 41 | 0.56 | 508 | 119 |
| 2003 | 36 | 0.84 | 767 | 138 |
| 2004 | 37 | 0.69 | 630 | 141 |
| 2005 | 54 | 0.78 | 710 | 143 |
| 2006 | 43 | 0.91 | 827 | 125 |
| 2007 | 49 | 0.76 | 692 | 132 |
| 2008 | 38 | 0.97 | 881 | 297 |
| 2009 | 45 | 0.80 | 732 | 142 |
| 2010 | 39 | 0.75 | 682 | 147 |
| 2011 | 45 | 0.58 | 533 | 87 |
| 2012 | 66 | 0.57 | 522 | 64 |
| 2013 | 51 | 0.73 | 668 | 125 |
| 2014 | 51 | 0.47 | 428 | 80 |
| 2015 | 51 | 0.73 | 664 | 127 |

Table 4．6．5．Nephrops，Firth of Forth（FU 8）：Summary of TV results for most recent 3 years（2013－ 2015）showing strata surveyed，numbers of stations in each strata，mean density and observed var－ iance，overall abundance and variance raised to stratum area．Proportion indicates relative amounts of overall raised variance attributable to each stratum．

|  | Area | Number OF | MEAN BURROW |  | Abundance | Stratum | Proportion |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | （ $\mathrm{km}^{2}$ ） | Stations | density | Observed | （millions） | variance | of total |
| Stratum |  |  | （no．／m²） | VARIANCE |  |  | variance |
| 2013 TV survey |  |  |  |  |  |  |  |
| M \＆SM | 170 | 10 | 0.477 | 0.342 | 81 | 992 | 0.213 |
| MS（west） | 139 | 8 | 0.568 | 0.214 | 79 | 518 | 0.111 |
| MS（mid） | 211 | 12 | 1.051 | 0.381 | 221 | 1409 | 0.302 |
| MS（east） | 395 | 21 | 0.725 | 0.235 | 286 | 1744 | 0.374 |
| Total | 915 | 51 |  |  | 668 | 4663 | 1 |
| 2014 TV survey |  |  |  |  |  |  |  |
| M \＆SM | 170 | 10 | 0.317 | 0.081 | 54 | 236 | 0.147 |
| MS（west） | 139 | 7 | 0.198 | 0.010 | 28 | 27 | 0.017 |
| MS（mid） | 211 | 12 | 0.725 | 0.134 | 153 | 496 | 0.309 |
| MS（east） | 395 | 22 | 0.491 | 0.119 | 194 | 847 | 0.527 |
| Total | 915 | 51 |  |  | 428 | 1606 | 1 |
| 2015 TV survey |  |  |  |  |  |  |  |
| M \＆SM | 170 | 9 | 0.613 | 0.447 | 105 | 1444 | 0.357 |
| MS（west） | 139 | 8 | 0.462 | 0.200 | 64 | 482 | 0.119 |
| MS（mid） | 211 | 12 | 0.955 | 0.243 | 201 | 898 | 0.222 |
| MS（east） | 395 | 22 | 0.746 | 0.173 | 295 | 1226 | 0.303 |
| Total | 915 | 51 |  |  |  | 4050 | 1 |

Table 4．6．6 Nephrops，Firth of Forth（FU 8）：Adjusted TV survey abundance，landings，total discard rate（proportion by number），dead discard rate and estimated harvest ratio 2003－2015．

| $\underset{\sim}{\underset{\sim}{u}}$ | 安気 | 【 岂 | $\leq \overline{\bar{z}}$ | 足 | щ ${ }_{\text {¢ }}$ ㅇ | ¢ $\bar{z}$ へ̌ | 気 | $\underline{\underline{0}}$ | ※ そ | ※ ¢ | 耑 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 767 | 12.4 | 51 | 59 | 95 | 1127 | 546 | 53.9 | 22.31 | 9.25 | 46.7 |
| 2004 | 630 | 16.4 | 74 | 40 | 103 | 1657 | 406 | 34.9 | 22.45 | 10.25 | 28.7 |
| 2005 | 710 | 19.4 | 89 | 65 | 138 | 1989 | 602 | 42.1 | 22.33 | 9.28 | 35.3 |
| 2006 | 827 | 26.7 | 115 | 142 | 221 | 2458 | 1510 | 55.2 | 21.43 | 10.67 | 48.1 |
| 2007 | 692 | 22.9 | 126 | 43 | 159 | 2651 | 614 | 25.3 | 20.97 | 14.34 | 20.3 |
| 2008 | 881 | 21.1 | 142 | 58 | 186 | 2450 | 796 | 29.1 | 17.23 | 13.65 | 23.5 |
| 2009 | 732 | 26 | 137 | 71 | 190 | 2663 | 573 | 34.1 | 19.41 | 8.09 | 27.9 |
| 2010 | 682 | 19.2 | 99 | 43 | 131 | 1950 | 407 | 30.2 | 19.76 | 9.55 | 24.5 |
| 2011 | 533 | 22.1 | 100 | 24 | 118 | 1889 | 231 | 19.5 | 19.75 | 9.56 | 15.3 |
| 2012 | 522 | 24.6 | 100 | 38 | 129 | 2129 | 379 | 27.2 | 21.66 | 10.10 | 21.9 |
| 2013 | 668 | 15.6 | 81 | 31 | 104 | 1501 | 301 | 27.4 | 19.30 | 9.82 | 22.0 |
| 2014 | 428 | 29.1 | 102 | 30 | 124 | 2382 | 353 | 22.9 | 24.30 | 11.66 | 18.3 |
| 2015 | 664 | 16.8 | 90 | 29 | 112 | 1892 | 311 | 24.4 | 21.84 | 10.74 | 19.5 |

Table 4.7.1 Nephrops, Moray Firth (FU 9), Nominal Landings (tonnes) of Nephrops, 1981-2015, as reported to the WG.

|  | UK SCOTLAND |  |  |  |  | Total ** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Nephrops trawl | Other <br> trawl | Creel | Sub-total |  |  |
|  |  |  |  |  |  |  |
| 1981 | 1299 | 117 | 0 | 1416 | 0 | 1416 |
| 1982 | 1033 | 86 | 0 | 1119 | 0 | 1119 |
| 1983 | 850 | 91 | 0 | 941 | 0 | 941 |
| 1984 | 960 | 209 | 0 | 1169 | 0 | 1169 |
| 1985 | 1908 | 173 | 0 | 2081 | 0 | 2081 |
| 1986 | 1932 | 211 | 0 | 2143 | 0 | 2143 |
| 1987 | 1724 | 268 | 0 | 1992 | 0 | 1992 |
| 1988 | 1637 | 322 | 0 | 1959 | 0 | 1959 |
| 1989 | 2102 | 474 | 0 | 2576 | 0 | 2576 |
| 1990 | 1698 | 339 | 0 | 2037 | 0 | 2037 |
| 1991 | 1285 | 235 | 0 | 1520 | 0 | 1520 |
| 1992 | 1285 | 306 | 0 | 1591 | 0 | 1591 |
| 1993 | 1505 | 304 | 0 | 1809 | 0 | 1809 |
| 1994 | 1179 | 358 | 0 | 1537 | 0 | 1537 |
| 1995 | 967 | 312 | 0 | 1279 | 0 | 1279 |
| 1996 | 1084 | 364 | 1 | 1449 | 2 | 1451 |
| 1997 | 1103 | 343 | 0 | 1446 | 1 | 1447 |
| 1998 | 739 | 289 | 4 | 1032 | 0 | 1032 |
| 1999 | 813 | 194 | 2 | 1009 | 0 | 1009 |
| 2000 | 1341 | 196 | 2 | 1539 | 0 | 1539 |
| 2001 | 1186 | 213 | 2 | 1401 | 0 | 1401 |
| 2002 | 883 | 247 | 2 | 1132 | 0 | 1132 |
| 2003 | 873 | 196 | 11 | 1080 | 0 | 1080 |
| 2004 | 1222 | 103 | 8 | 1333 | 0 | 1333 |
| 2005 | 1526 | 64 | 12 | 1602 | 3 | 1605 |
| 2006 | 1751 | 42 | 11 | 1804 | 1 | 1805 |
| 2007 | 1818 | 17 | 6 | 1841 | 2 | 1843 |
| 2008 | 1444 | 68 | 3 | 1515 | 0 | 1515 |
| 2009 | 1033 | 31 | 2 | 1066 | 1 | 1067 |
| 2010 | 1026 | 28 | 9 | 1063 | 0 | 1063 |
| 2011 | 1358 | 23 | 9 | 1390 | 1 | 1391 |
| 2012 | 834 | 24 | 8 | 866 | 0 | 866 |
| 2013 | 497 | 116 | 7 | 620 | 3 | 623 |
| 2014 | 890 | 348 | 2 | 1240 | 12 | 1252 |
| 2015* | 604 | 224 | 0 | 828 | 2 | 830 |
| * provisional na = not available |  |  |  |  |  |  |

** No landings by non UK countries from this FU

Table 4.7.2 Nephrops, Moray Firth (FU 9): landings, effort (days fishing) and LPUE (kg/day) for UK bottom trawlers landing in Scotland and fishing Nephrops with codend mesh sizes of 70 mm or above, 2000-2015.

|  | YeAr | LANDINGS (TONNES) | Effort (DAYS) |
| :--- | :--- | :--- | :--- |
| 2000 | 1537 | 7943 | LPUE (KG/DAY) |
| 2001 | 1399 | 7219 | 193.5 |
| 2002 | 1130 | 7495 | 193.8 |
| 2003 | 1069 | 5934 | 150.8 |
| 2004 | 1325 | 6200 | 180.1 |
| 2005 | 1590 | 4805 | 213.7 |
| 2006 | 1793 | 4588 | 330.9 |
| 2007 | 1835 | 4758 | 390.8 |
| 2008 | 1512 | 4328 | 385.7 |
| 2009 | 1064 | 3546 | 349.4 |
| 2010 | 1054 | 3589 | 300.1 |
| 2011 | 1381 | 3880 | 293.7 |
| 2012 | 858 | 3079 | 355.9 |
| 2013 | 613 | 2954 | 278.7 |
| 2014 | 1238 | 4099 | 207.5 |
| $2015^{*}$ | 828 | 3755 | 302.0 |

* provisional na = not available

Table 4.7.3 Nephrops, Moray Firth (FU 9): Mean sizes (CL mm) above and below 35 mm of male and female Nephrops in Scottish catches and landings, 1991-2015.

| Year | Catches |  | Landings |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | < 35 mm CL |  | < 35 mm CL |  | => 35 mm CL |  |
|  | Males | Females | Males | Females | Males | Females |
| 1981 | na | na | 30.5 | 28.2 | 39.1 | 37.7 |
| 1982 | na | na | 30.2 | 29.0 | 40.0 | 37.9 |
| 1983 | na | na | 29.9 | 29.1 | 40.6 | 38.3 |
| 1984 | na | na | 29.7 | 29.3 | 39.4 | 38.1 |
| 1985 | na | na | 28.9 | 28.7 | 38.7 | 37.8 |
| 1986 | na | na | 28.7 | 27.8 | 39.1 | 38.4 |
| 1987 | na | na | 29.0 | 28.3 | 39.4 | 38.6 |
| 1988 | na | na | 29.1 | 28.7 | 38.9 | 38.4 |
| 1989 | na | na | 29.8 | 28.8 | 40.1 | 39.4 |
| 1990 | 28.8 | 28.1 | 30.3 | 29.1 | 38.4 | 38.7 |
| 1991 | 28.3 | 27.4 | 30.1 | 28.6 | 38.2 | 38.2 |
| 1992 | 29.4 | 28.6 | 31.0 | 30.5 | 38.3 | 38.0 |
| 1993 | 29.8 | 29.9 | 31.3 | 30.9 | 38.6 | 37.7 |
| 1994 | 28.9 | 30.1 | 30.8 | 31.0 | 39.4 | 37.5 |
| 1995 | 25.8 | 25.0 | 29.9 | 29.3 | 39.1 | 38.0 |
| 1996 | 29.3 | 28.4 | 30.6 | 29.7 | 38.5 | 38.0 |
| 1997 | 28.5 | 27.9 | 29.5 | 28.9 | 38.8 | 38.2 |
| 1998 | 28.7 | 28.2 | 30.1 | 29.3 | 38.8 | 38.2 |
| 1999 | 29.5 | 28.8 | 30.4 | 29.7 | 38.9 | 37.6 |
| 2000 | 29.8 | 29.1 | 31.5 | 30.6 | 39.2 | 38.3 |
| 2001 | 30.0 | 29.2 | 30.9 | 30.2 | 39.5 | 37.9 |
| 2002 | 27.2 | 27.0 | 31.2 | 30.9 | 41.0 | 38.7 |
| 2003 | 29.3 | 29.2 | 30.3 | 30.1 | 39.8 | 38.0 |
| 2004 | 29.3 | 28.4 | 31.3 | 30.8 | 39.0 | 39.2 |
| 2005 | 30.0 | 28.7 | 31.0 | 29.6 | 39.2 | 38.5 |
| 2006 | 29.7 | 28.9 | 30.6 | 29.6 | 39.3 | 38.6 |
| 2007 | 30.1 | 28.8 | 30.3 | 29.0 | 39.4 | 38.6 |
| 2008 | 29.3 | 27.7 | 30.2 | 28.2 | 39.8 | 40.2 |
| 2009 | 29.7 | 28.9 | 30.7 | 29.3 | 39.6 | 38.5 |
| 2010 | 29.7 | 29.1 | 31.1 | 30.5 | 40.0 | 38.9 |
| 2011 | 28.6 | 28.4 | 29.4 | 29.0 | 39.5 | 38.4 |
| 2012 | 29.5 | 29.1 | 30.5 | 29.9 | 39.2 | 38.5 |
| 2013 | 30.7 | 29.3 | 30.9 | 29.5 | 39.6 | 38.4 |
| 2014 | 30.2 | 29.8 | 31.6 | 30.8 | 40.3 | 39.0 |
| 2015 | 29.8 | 29.4 | 31.5 | 30.6 | 40.6 | 39.1 |
| na = not available |  |  |  |  |  |  |

Table 4.7.4 Nephrops, Moray Firth (FU 9): Results of the 1993-2015 TV surveys.


Table 4.7.5 Nephrops, Moray Firth (FU 9):Summary of TV results for most recent 3 years (2013-2015) showing strata surveyed, numbers of stations in each strata, mean density and observed variance, overall abundance and variance raised to stratum area. Proportion indicates relative amounts of overall raised variance attributable to each stratum.

| Stratum |  | Number <br> OF | Mean BURROW | Observed variance | $\frac{\text { ABUNDANCE }}{\text { (millions) }}$ | Stratum | PROPORTION <br> of total <br> variance |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\frac{\text { AREA }}{\left(\mathrm{km}^{2}\right)}$ | Stations | density |  |  |  |  |
|  |  |  | (no./m²) |  |  |  |  |
| 2013 TV surve |  |  |  |  |  |  |  |
| M \& SM | 169 | 3 | 0.22 | 0.06 | 37 | 580 | 0.17 |
| MS(west) | 682 | 18 | 0.22 | 0.06 | 148 | 1576 | 0.463 |
| MS(mid) | 698 | 18 | 0.23 | 0.01 | 160 | 300 | 0.088 |
| MS(east) | 646 | 16 | 0.19 | 0.04 | 124 | 950 | 0.279 |
| Total | 2195 | 55 |  |  | 469 | 3406 | 1 |
| 2014 TV surve |  |  |  |  |  |  |  |
| M \& SM | 169 | 3 | 0.19 | 0.04 | 33 | 412 | 0.202 |
| MS(west) | 682 | 16 | 0.14 | 0.03 | 98 | 851 | 0.417 |
| MS(mid) | 698 | 17 | 0.15 | 0.02 | 103 | 436 | 0.213 |
| MS(east) | 646 | 16 | 0.15 | 0.01 | 97 | 344 | 0.168 |
| Total | 2195 | 52 |  |  | 331 | 2042 | 1 |
| 2015 TV surve |  |  |  |  |  |  |  |
| M \& SM | 169 | 3 | 0.30 | 0.03 | 51 | 235 | 0.134 |
| MS(west) | 682 | 19 | 0.11 | 0.02 | 75 | 542 | 0.309 |
| MS(mid) | 698 | 17 | 0.22 | 0.02 | 151 | 456 | 0.259 |
| MS(east) | 646 | 13 | 0.11 | 0.02 | 71 | 525 | 0.299 |
| Total | 2195 | 52 |  |  | 347 | 1757 | 1 |

Table 4.7.6 Nephrops, Moray Firth (FU 9): Adjusted TV survey abundance, landings, discard rate (proportion by number), dead discard rate (proportion by number) and estimated harvest ratio 2003-2015.

| $\stackrel{\underset{\sim}{\underset{\sim}{\underset{~}{4}}}}{ }$ |  |  |  |  |  |  |  | $\begin{aligned} & \hline \stackrel{\mu}{k} \\ & \stackrel{y}{\kappa} \\ & 0 \\ & \stackrel{y}{c} \\ & \frac{n}{0} \end{aligned}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 730 | 7.1 | 46 | 7 | 52 | 1080 | 70 | 13.7 | 23.32 | 9.51 | 10.6 |
| 2004 | 626 | 10.5 | 48 | 23 | 66 | 1333 | 272 | 32.6 | 27.57 | 11.62 | 26.6 |
| 2005 | 869 | 8.8 | 67 | 12 | 76 | 1605 | 122 | 15.0 | 23.84 | 10.31 | 11.7 |
| 2006 | 445 | 20.1 | 81 | 12 | 90 | 1805 | 117 | 12.8 | 22.34 | 9.86 | 9.9 |
| 2007 | 531 | 16 | 80 | 7 | 85 | 1843 | 95 | 7.9 | 23.04 | 13.95 | 6.0 |
| 2008 | 481 | 13.7 | 60 | 8 | 66 | 1515 | 74 | 11.4 | 25.29 | 9.60 | 8.8 |
| 2009 | 415 | 11.6 | 45 | 4 | 48 | 1067 | 33 | 7.6 | 23.46 | 8.72 | 5.8 |
| 2010 | 406 | 11.5 | 39 | 10 | 47 | 1063 | 104 | 19.8 | 26.94 | 10.63 | 15.7 |
| 2011 | 372 | 18.9 | 63 | 10 | 70 | 1391 | 102 | 13.9 | 21.63 | 10.12 | 10.8 |
| 2012 | 299 | 13.7 | 37 | 6 | 41 | 866 | 54 | 13.2 | 23.16 | 9.72 | 10.3 |
| 2013 | 469 | 5.8 | 26 | 1 | 27 | 655 | 10 | 3.3 | 24.95 | 11.21 | 2.5 |
| 2014 | 331 | 14.7 | 43 | 7 | 49 | 1252 | 87 | 14.6 | 28.94 | 11.79 | 11.3 |
| 2015 | 347 | 9.1 | 28 | 5 | 32 | 830 | 56 | 15.1 | 29.1 | 11.35 | 11.8 |

Table 4.8.1 Nephrops, Noup (FU 10), Nominal Landings (tonnes) of Nephrops, 1981-2015, as reported to the WG.

| Year | NEPHROPS TRAWL | Other trawl | Creel | Sub Total | Other UK | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 12 | 23 | 0 | 35 | 0 | 35 |
| 1982 | 12 | 7 | 0 | 19 | 0 | 19 |
| 1983 | 10 | 6 | 0 | 16 | 0 | 16 |
| 1984 | 76 | 35 | 0 | 111 | 0 | 111 |
| 1985 | 1 | 21 | 0 | 22 | 0 | 22 |
| 1986 | 45 | 22 | 0 | 67 | 0 | 67 |
| 1987 | 13 | 32 | 0 | 45 | 0 | 45 |
| 1988 | 23 | 53 | 0 | 76 | 0 | 76 |
| 1989 | 24 | 60 | 0 | 84 | 0 | 84 |
| 1990 | 101 | 117 | 0 | 218 | 0 | 218 |
| 1991 | 111 | 86 | 0 | 197 | 0 | 197 |
| 1992 | 58 | 130 | 0 | 188 | 0 | 188 |
| 1993 | 200 | 176 | 0 | 376 | 0 | 376 |
| 1994 | 307 | 187 | 0 | 494 | 0 | 494 |
| 1995 | 163 | 116 | 0 | 279 | 0 | 279 |
| 1996 | 181 | 164 | 0 | 345 | 0 | 345 |
| 1997 | 185 | 131 | 1 | 317 | 0 | 317 |
| 1998 | 184 | 72 | 0 | 256 | 0 | 256 |
| 1999 | 211 | 67 | 0 | 278 | 0 | 278 |
| 2000 | 196 | 78 | 0 | 274 | 0 | 274 |
| 2001 | 88 | 89 | 0 | 177 | 0 | 177 |
| 2002 | 246 | 157 | 0 | 403 | 0 | 403 |
| 2003 | 258 | 78 | 0 | 336 | 0 | 336 |
| 2004 | 174 | 54 | 0 | 228 | 0 | 228 |
| 2005 | 81 | 84 | 0 | 165 | 0 | 165 |
| 2006 | 44 | 89 | 0 | 133 | 0 | 133 |
| 2007 | 46 | 107 | 0 | 153 | 0 | 153 |
| 2008 | 74 | 98 | 0 | 172 | 0 | 172 |
| 2009 | 24 | 63 | 0 | 87 | 0 | 87 |
| 2010 | 4 | 35 | 0 | 39 | 0 | 39 |
| 2011 | 27 | 41 | 0 | 68 | 0 | 68 |
| 2012 | 2 | 11 | 0 | 13 | 0 | 13 |
| 2013 | 4 | 12 | 0 | 16 | 0 | 16 |
| 2014 | 5 | 9 | 1 | 15 | 0 | 15 |
| 2015 | 5 | 10 | 0 | 15 | 0 | 15 |

* provisional

Table 4.8.2 Nephrops, Noup (FU 10): landings (tonnes), effort (days fishing) and LPUE (kg/day) for UK bottom trawlers landing in Scotland and fishing Nephrops with codend mesh sizes of 70 mm or above, 2000-2015.

|  | YeAR | LANDINGS (TONNES) | EfFORT (DAYS) |
| :--- | :--- | :--- | :--- |
| 2000 | 274 | 1622 | LPUE (KG/DAY) |
| 2001 | 177 | 1383 | 168.9 |
| 2002 | 403 | 2036 | 128.0 |
| 2003 | 336 | 1434 | 197.9 |
| 2004 | 228 | 899 | 234.3 |
| 2005 | 165 | 730 | 253.6 |
| 2006 | 133 | 612 | 226.0 |
| 2007 | 153 | 591 | 217.3 |
| 2008 | 172 | 746 | 258.9 |
| 2009 | 87 | 871 | 230.6 |
| 2010 | 39 | 813 | 99.9 |
| 2011 | 68 | 776 | 48.0 |
| 2012 | 13 | 574 | 87.6 |
| 2013 | 16 | 454 | 22.6 |
| 2014 | 14 | 673 | 35.2 |
| $2015^{*}$ | 15 | 514 | 20.8 |
| * provisional |  |  | 29.2 |

Table 4.8.3 Nephrops, Noup (FU 10): Mean sizes (CL mm) above and below 35 mm of male and female Nephrops in landings, 1997-2015. No females in samples in 2010 and no sampling in 2015.

| Year | Landings |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | < 35 mm CL |  | $\Rightarrow 35 \mathrm{~mm} \mathrm{CL}$ |  |
|  | Males | Females | Males | Females |
| 1997 | 29.7 | 28.3 | 40.4 | 38.2 |
| 1998 | 30.4 | 29.8 | 38.8 | 38.6 |
| 1999 | 30.4 | 30.1 | 39.2 | 37.8 |
| 2000 | 31.8 | 30.1 | 38.2 | 39.1 |
| 2001 | 31.4 | 29.5 | 38.7 | 37.9 |
| 2002 | 30.8 | 29.9 | 39.7 | 38.5 |
| 2003 | 29.3 | 30.4 | 39.9 | 38.5 |
| 2004 | 31.4 | 30.0 | 40.2 | 38.8 |
| 2005 | 31.0 | 29.3 | 39.3 | 38.4 |
| 2006 | 30.8 | 30.2 | 40.4 | 38.7 |
| 2007 | 30.7 | 29.4 | 40.2 | 38.7 |
| 2008 | 31.9 | 30.6 | 40.3 | 39.3 |
| 2009 | 33.2 | 33.2 | 42.6 | 42.7 |
| 2010 | 33.3 | na | 42.6 | na |
| 2011 | 32.8 | 32.7 | 43.3 | 40.1 |
| 2012 | 32.4 | 31.8 | 40.7 | 40.1 |
| 2013 | 34.0 | 32.4 | 43.7 | 39.7 |
| 2014 | 33.3 | 33.0 | 46.6 | 43.2 |
| 2015 | na | na | na | na |
| na = not available |  |  |  |  |

Table 4.8.4 Nephrops, Noup (FU 10): Results of the 1994, 1999, 2006, 2007 \& 2014 TV surveys (absolute conversion factor $=\mathbf{1 . 3 5}$, from Fladen).


Table 4.9.1 Nephrops Norwegian Deep (FU 32). Landings (tonnes) by country, 1993-2015, estimated Danish discards (2003-2015), and TAC (EU). The 2005 discards numbers were updated in 2015.

| Year | Denmark | Danish discards |  | Norway |  |  | Sweden | UK | Netherlands | Total | TAC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | dead | live | Trawl | Creel | Sub-total |  |  |  |  |  |
| 1993 | 220 |  |  | 102 | 1 | 103 |  | 16 |  | 339 |  |
| 1994 | 584 |  |  | 161 | 0 | 161 |  | 10 |  | 755 |  |
| 1995 | 418 |  |  | 68 | 1 | 69 |  | 2 |  | 489 |  |
| 1996 | 868 |  |  | 73 | 1 | 74 |  | 10 |  | 952 |  |
| 1997 | 689 |  |  | 56 | 8 | 64 |  | 7 |  | 760 |  |
| 1998 | 743 |  |  | 88 | 1 | 89 |  | 4 |  | 836 |  |
| 1999 | 972 |  |  | 119 | 15 | 134 |  | 13 |  | 1119 |  |
| 2000 | 871 |  |  | 143 | 0 | 143 | 37 | 34 |  | 1085 |  |
| 2001 | 1026 |  |  | 72 | 13 | 85 | 26 | 53 |  | 1190 |  |
| 2002 | 1043 |  |  | 42 | 21 | 63 | 13 | 52 |  | 1171 |  |
| 2003 | 996 | 145 | 48 | 68 | 11 | 79 | 1 | 14 |  | 1090 |  |
| 2004 | 835 | 200 | 67 | 72 | 8 | 80 | 1 | 6 |  | 922 | 1000 |
| 2005 | 979 | 194 | 65 | 89 | 13 | 102 | 2 | 6 |  | 1089 | 1000 |
| 2006 | 939 | 126 | 42 | 62 | 19 | 81 | 1 | 7 | 5 | 1033 | 1300 |
| 2007 | 652 | 64 | 21 | 77 | 20 | 97 | 5 | 1 |  | 755 | 1300 |
| 2008 | 505 |  |  | 112 | 30 | 142 | 24 | 4 |  | 675 | 1300 |
| 2009 | 331 | 29 | 10 | 107 | 31 | 138 | 2 | 6 |  | 477 | 1200 |
| 2010 | 282 | 36 | 12 | 82 | 41 | 123 | 1 | 1 |  | 407 | 1200 |
| 2011 | 322 |  |  | 29 | 40 | 69 | 1 | 3 |  | 395 | 1200 |
| 2012 | 234 | 35 | 12 | 25 | 50 | 75 | 1 | 0 |  | 310 | 1200 |
| 2013 | 128 | 51 | 17 | 18 | 45 | 63 | 0 | 0 |  | 191 | 1000 |
| 2014 | 143 | 4 | 1 | 15 | 47 | 62 | 0 | 0 |  | 205 | 1000 |
| 2015* | 110 | 5 | 2 | 8 | 74 | 82 | 0 | 0 |  | 192 | 1000 |

* provisional

Table 4.9.2 Nephrops Norwegian Deep (FU 32). Danish effort (kW days, days at sea, fishing days) and LPUE (kg/kW day) for bottom trawlers catching Nephrops, 1993-2015. Effort values were updated in 2016.

| Year | $\begin{aligned} & \text { KW DAYS } \\ & (' 1000) \end{aligned}$ | DAYS AT SEA | Fishing days | LPUE |
| :---: | :---: | :---: | :---: | :---: |
| 1993 | 891 | 1980 | 1536 | 247 |
| 1994 | 1439 | 3574 | 2793 | 406 |
| 1995 | 1009 | 2464 | 1936 | 414 |
| 1996 | 1734 | 4000 | 3229 | 501 |
| 1997 | 1962 | 4162 | 3410 | 351 |
| 1998 | 1471 | 3251 | 2644 | 505 |
| 1999 | 2262 | 4658 | 3763 | 430 |
| 2000 | 2662 | 5068 | 4152 | 327 |
| 2001 | 3511 | 6429 | 5464 | 292 |
| 2002 | 3105 | 5743 | 4791 | 336 |
| 2003 | 3494 | 6287 | 5404 | 285 |
| 2004 | 2443 | 4297 | 3653 | 342 |
| 2005 | 2787 | 5076 | 4348 | 351 |
| 2006 | 3023 | 5274 | 4514 | 311 |
| 2007 | 1782 | 3052 | 2557 | 366 |
| 2008 | 1589 | 2521 | 2123 | 318 |
| 2009 | 1351 | 2160 | 1793 | 245 |
| 2010 | 1151 | 1903 | 1612 | 245 |
| 2011 | 1152 | 1863 | 1543 | 280 |
| 2012 | 907 | 1474 | 1224 | 258 |
| 2013 | 862 | 1450 | 1200 | 149 |
| 2014 | 747 | 1224 | 1054 | 191 |
| 2015 | 576 | 927 | 784 | 191 |

Table 4.10.1 Nephrops in FU 33. (Off Horns Reef) Landings (tonnes) by country, 1993-2013.

|  | Belgium | Denmark | Germany | Netherl. | UK | Total ** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 159 |  | na | 1 | 160 |
| 1994 | 0 | 137 |  | na | 0 | 137 |
| 1995 | 3 | 158 |  | 3 | 1 | 164 |
| 1996 | 1 | 74 |  | 2 | 0 | 77 |
| 1997 | 0 | 274 |  | 2 | 0 | 276 |
| 1998 | 4 | 333 | 8 | 12 | 1 | 350 |
| 1999 | 22 | 683 | 14 | 12 | 6 | 724 |
| 2000 | 13 | 537 | 12 | 39 | 9 | 597 |
| 2001 | 52 | 667 | 11 | 61 | + | 791 |
| 2002 | 21 | 772 | 13 | 51 | 4 | 861 |
| 2003 | 15 | 842 | 4 | 67 | 1 | 929 |
| 2004 | 37 | 1097 | 24 | 109 | 1 | 1268 |
| 2005 | 16 | 803 | 31 | 191 | 9 | 1050 |
| 2006 | 97 | 710 | 151 | 314 | 15 | 1288 |
| 2007 | 118 | 610 | 201 | 496 | 42 | 1467 |
| 2008 | 130 | 362 | 160 | 386 | 58 | 1096 |
| 2009 | 121 | 231 | 150 | 491 | 170 | 1163 |
| 2010 | 56 | 180 | 206 | 295 | 69 | 806 |
| 2011 | 163 | 396 | 202 | 403 | 28 | 1191 |
| 2012 | 181 | 394 | 132 | 376 | 2 | 1084 |
| 2013 | 156 | 310 | 174 | 304 | 2 | 946 |
| 2014 | 229 | 387 | 161 | 360 | 9 | 1146 |
| 2015* | 299 | 371 | 142 | 187 | 4 | 1003 |

[^3]Table 4.10.2 Nephrops in FU 33. (Off Horns Reef): Danish logbook recorded effort (kW days, Days at sea and fishing days) and LPUE (kg/kW day) for bottom trawlers catching Nephrops with codend mesh sizes of 70 mm or above, 1991-2015.

| YeAR | KW DAYS | DAYS AT SEA | FISHING DAYS | LPUE* |
| :--- | :--- | :--- | :--- | :--- |
| 1991 | 596367.7 | 1363 | 1087 | 0.12 |
| 1992 | 533565 | 1382 | 1068 | 0.14 |
| 1993 | 628812.5 | 1441 | 1141 | 0.25 |
| 1994 | 387571.7 | 997 | 782 | 0.35 |
| 1995 | 376068.1 | 1068 | 813 | 0.42 |
| 1996 | 212737.4 | 634 | 494 | 0.35 |
| 1997 | 490267.8 | 1446 | 1126 | 0.56 |
| 1998 | 752999.4 | 2254 | 1741 | 0.44 |
| 1999 | 1168914 | 3400 | 2714 | 0.58 |
| 2000 | 1039983 | 3200 | 2473 | 0.52 |
| 2001 | 1251480 | 3836 | 3049 | 0.53 |
| 2002 | 1610003 | 4544 | 3533 | 0.48 |
| 2003 | 1598038 | 4722 | 3795 | 0.53 |
| 2004 | 1900555 | 5626 | 4407 | 0.58 |
| 2005 | 1084823 | 3276 | 2624 | 0.74 |
| 2006 | 959737.6 | 2703 | 2146 | 0.74 |
| 2007 | 773976.6 | 1972 | 1548 | 0.79 |
| 2008 | 445158.7 | 926 | 722 | 0.81 |
| 2009 | 274715.9 | 647 | 547 | 0.84 |
| 2010 | 246931.1 | 528 | 425 | 0.73 |
| 2011 | 346294.2 | 760 | 608 | 1.14 |
| 2012 | 298139 | 700 | 589 | 1.32 |
| 2013 | 238654.1 | 560 | 492 | 1.30 |
| 2014 | 374372.2 | 882 | 752 | 1.03 |
| 2015 | 279017.5 | 663 | 586 | 1.33 |
| $k W$ |  |  |  |  |

*kg/ kW days

Table 4.11.1. Nephrops, Devil's Hole (FU 34). Nominal landings (tonnes) of Nephrops 1986-2015 as reported to the WG. Scottish data only from 1986 to 2009.

| Year | UK Scotland |  |  |  |  | Denmark | Netherlands | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Nephrops trawl | Other trawl | Creel | Sub-total |  |  |  |  |
| 1986 | 20 | 3 | 0 | 23 |  |  |  | 23 |
| 1987 | 2 | 3 | 0 | 5 |  |  |  | 5 |
| 1988 | 1 | 1 | 0 | 2 |  |  |  | 2 |
| 1989 | 15 | 13 | 0 | 28 |  |  |  | 28 |
| 1990 | 20 | 6 | 0 | 26 |  |  |  | 26 |
| 1991 | 64 | 21 | 0 | 85 |  |  |  | 85 |
| 1992 | 78 | 28 | 0 | 106 |  |  |  | 106 |
| 1993 | 23 | 21 | 0 | 44 |  |  |  | 44 |
| 1994 | 79 | 50 | 0 | 129 |  |  |  | 129 |
| 1995 | 37 | 95 | 0 | 132 |  |  |  | 132 |
| 1996 | 40 | 89 | 0 | 129 |  |  |  | 129 |
| 1997 | 30 | 70 | 0 | 100 |  |  |  | 100 |
| 1998 | 15 | 73 | 0 | 88 |  |  |  | 88 |
| 1999 | 80 | 122 | 0 | 202 |  |  |  | 202 |
| 2000 | 89 | 95 | 0 | 184 |  |  |  | 184 |
| 2001 | 159 | 112 | 0 | 271 |  |  |  | 271 |
| 2002 | 240 | 103 | 0 | 343 |  |  |  | 343 |
| 2003 | 518 | 157 | 0 | 675 |  |  |  | 675 |
| 2004 | 398 | 90 | 0 | 488 |  |  |  | 488 |
| 2005 | 253 | 125 | 0 | 378 |  |  |  | 378 |
| 2006 | 359 | 89 | 0 | 448 |  |  |  | 448 |
| 2007 | 649 | 68 | 0 | 717 |  |  |  | 717 |
| 2008 | 844 | 93 | 0 | 937 |  |  |  | 937 |
| 2009 | 1297 | 8 | 0 | 1305 |  |  |  | 1305 |
| 2010 | 816 | 22 | 0 | 838 | 25 | 1 | 1 | 865 |
| 2011 | 406 | 16 | 0 | 422 | 6 | 4 |  | 432 |
| 2012 | 546 | 4 | 0 | 550 | 37 | 10 |  | 597 |
| 2013 | 65 | 41 | 0 | 106 | 11 | 3 |  | 120 |
| 2014 | 81 | 226 | 0 | 307 | 13 |  |  | 320 |
| 2015* | 218 | 182 | 0 | 400 | 39 | <0.5 |  | 439 |

[^4]Table 4.11.2 Nephrops, Devils Hole (FU 34): landings, effort (days fishing) and LPUE (kg/day) for UK bottom trawlers landing in Scotland and fishing Nephrops with codend mesh sizes of 70 mm or above, 2000-2015.

|  | Year | LANDINGS (ToNNES) | Effort (DAYS) |
| :--- | :--- | :--- | :--- |
| 2000 | 184 | 3391 | LPUE (KG/DAY) |
| 2001 | 271 | 3142 | 54.3 |
| 2002 | 343 | 2022 | 86.3 |
| 2003 | 675 | 2614 | 169.6 |
| 2004 | 488 | 1551 | 258.2 |
| 2005 | 378 | 1545 | 314.6 |
| 2006 | 448 | 1440 | 244.7 |
| 2007 | 717 | 1824 | 311.1 |
| 2008 | 937 | 1673 | 393.1 |
| 2009 | 1305 | 1921 | 560.1 |
| 2010 | 838 | 1465 | 679.3 |
| 2011 | 422 | 1041 | 572.0 |
| 2012 | 550 | 1255 | 405.4 |
| 2013 | 106 | 438 | 438.2 |
| 2014 | 307 | 758 | 242.0 |
| 2015 | 400 | 1222 | 405.0 |

* provisional

Table 4.11.3. Nephrops, Devil's Hole (FU 34). Mean sizes (CL mm ) above and below 35 mm of male and female Nephrops in Scottish catches and landings, 2006-2015. Samples not available in 2012 and 2013.

|  | Landings <br>  <br>  <br>  <br> YEAR |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
|  | Males | Females | Males | Females |
| 2006 | 29.7 | 29.8 | 39.7 | 38.1 |
| 2007 | 30.4 | 28.7 | 40.5 | 39.2 |
| 2008 | 31 | 30.5 | 40.3 | 39.6 |
| 2009 | 31.7 | 31.1 | 41.3 | 40.6 |
| 2010 | 32.1 | 29.7 | 39.1 | 38.8 |
| 2011 | 31.7 | 30.7 | 43.7 | 40.4 |
| 2012 | na | na | na | na |
| 2013 | na | na | na | na |
| 2014 | 33.0 | 34.0 | 42.0 | 41.4 |
| 2015 | na | na | na | na |
|  | na $=$ not available |  |  |  |

Table 4.11.4. Nephrops, Devil's Hole (FU 34). Results of the 2003, 2005, 2009-12 and 2014-2015 surveys.

| Year | Stations | Mean | 95\% |
| :---: | :---: | :---: | :---: |
|  |  | density | confidence |
|  |  |  | interval |
|  |  | burrows/m ${ }^{2}$ | burrows/m ${ }^{2}$ |
| 2003 | 20 | 0.09 | 0.02 |
| 2004 | no survey |  |  |
| 2005 | 29 | 0.09 | 0.04 |
| 2006 | no survey |  |  |
| 2007 | no survey |  |  |
| 2008 | no survey |  |  |
| 2009 | 12 | 0.28 | 0.13 |
| 2010 | 19 | 0.24 | 0.08 |
| 2011 | 14 | 0.16 | 0.09 |
| 2012 | 15 | 0.14 | 0.06 |
| 2013 | no survey |  |  |
| 2014 | 13 | 0.13 | 0.04 |
| 2015 | 17 | 0.16 | 0.06 |

## Length frequencies for catch (dotted) and landed(solid): Nephrops in FU 5



Figure 4.3.1 - FU5 Botney Gut/Silver Pit. Size distribution for Dutch landings, from 2004 to 2011.


Figure 4.3.2 FU5 Botney Gut/Silver Pit. Long-term trends in landings, effort and LPUE.


Figure 4.4.1 Nephrops in FU6. Landings, directed effort, directed LPUE and mean sizes of different catch components.


Figure 4.4.2 Nephrops in FU6, Number of participating vessels (from UK) by vessel size category.


Figure 4.4.3 Nephrops in FU6, annual discard ogives. The different point shapes represent different sampling trips within any year.

FU6: Quarterly Male Sex ratio


Figure 4.4.4 Nephrops in FU6: Quarterly sex ratio in the catches.


Figure 4.4.5 Nephrops in FU6: LPUE for directed English trawlers by gear type and vessel size




Figure 4.4.6 Nephrops in FU6: LPUE by sex and quarter.

## Length frequencies for catch (dotted) and landed(solid): Nephrops in fu6



Figure 4.4.7 Nephrops in FU6: Annual length frequencies for landings and discards.


Figure 4.4.8 Nephrops in FU6: Breakdown of landings by gear, country and sample availibilty.

FU6: Stock abundance


Figure 4.4.9 Nephrops in FU6: Time series of UWTV results. The dashed green line is the proxy for MSY Btriger , the abundance estimate for 2007. The red line since 2007 gives the Geostatsistical abundance estimate. Prior to 2007 the estimate was raised using stratified boxes of ground but due to the spatial distribution of stations was biased.


Figure 4.4.10 Nephrops in FU6: Results of the UWTV survey.


Figure 4.4.11 Nephrops in FU6: Observed harvest ratio (removals divided by abundance estimate).


Figure 4.4.12 Nephrops in FU6: Separable Cohort analysis model fit. Solid lines are for males, dashed lines are females, thick lines represent the landings component, the thin lines represent the discarded component. The top left panel gives observed and predicted numbers at length in the discards and landings, top right gives the fishing mortality at length with the vertical lines representing length at $25 \%$ selection and $50 \%$ selection. Bottom left shows residual numbers (observed - expected) at length. The bottom right gives the Yield Per recruit against fishing mortality, the thick solid line gives the combined value and vertical lines represent ${ }_{0.1}$ for the three curves.


LPUE - Scottish Nephrops trawlers



Figure 4.5.1 Nephrops, Fladen (FU 7), Long term landings, effort, LPUE and mean sizes. Note that the effort and LPUE from Scottish trawlers cover a shorter period 2000-2015.


Figure 4.5.2 Nephrops, Fladen (FU 7), Landings by quarter and sex from Scottish Nephrops trawlers.


Figure4.5.3 Nephrops Fladen Ground (FU 7)Length composition of catch of males (right) and females left from 2000 (bottom) to 2015 (top). Mean sizes of catch and landings are displayed vertically.

4.5.4 Nephrops, (FUs 7-9 and 34, Fladen, Firth of Forth, Moray Firth and Devil's Hole). Individual mean weight (g) in the landings from 1990-2015 (Scottish market sampling data). FU 34 data only shown for 2006-2011.


Figure 4.5.5 Nephrops, Fladen (FU 7). TV survey distribution and relative density (2010-2015). Green and brown areas represent areas of suitable sediment for Nephrops. Density proportional to circle radius. Red crosses represent zero observations.


Figure 4.5.6 Nephrops, Fladen (FU 7), Time series of TV survey abundance estimates with $95 \%$ confidence intervals, 1992-2015.


Figure 4.5.7 Nephrops, Fladen (FU 7), Quarterly sex ratio (by number) in catches.


Figure 4.5.8 Nephrops, Fladen (FU 7), VMS distribution of vessels in Fladen (2010-2015). Points in figure correspond to fishing pings (speed<5 kn) associated with trips made by otter trawlers landing more than $25 \%$ of Nephrops by weight.


Figure 4.5.9 Nephrops, Fladen (FU 7), UWTV density by sediment type in the North (left plot) and South (right plot) of Fladen (split at the 58.75 N latitude line). F: fine sediment (silt \& clay $>80 \%$ ); MF: medium fine sediment ( $55 \%<$ silt \& clay< 80 ); MC: medium coarse sediment $(40 \%<$ silt \& clay< 55); C: coarse sediment (silt \& clay $<40 \%$ ).

## Landings - International



LPUE - Scottish Nephrops trawlers



Figure 4.6.1 Nephrops, Firth of Forth (FU 8), Long term landings and mean sizes. Note that the effort and LPUE from Scottish trawlers cover a shorter period 2000-2015.


Figure 4.6.2 Nephrops, Firth of Forth (FU 8), Landings by quarter and sex from Scottish Nephrops trawlers.


Figure 4.6.3 Nephrops Firth of Forth (FU 8) Length composition of catch of males (right) and females left from 2000 (bottom) to 2015 (top). Mean sizes of catch and landings are displayed vertically.


Figure 4.6.4 Nephrops, Firth of Forth (FU 8). TV survey distribution and relative density (2010-2015). Green and brown areas represent areas of suitable sediment for Nephrops. Density proportional to circle radius. Red crosses represent zero observations.

## firth forth



Figure 4.6.5 Nephrops, Firth of Forth (FU 8), Time series of TV survey abundance estimates with 95\% confidence intervals, 1993-2015.


Figure 4.6.6 Nephrops, Firth of Forth (FU 8), Quarterly sex ratio (by number) in catches.

## Landings - Internationa



LPUE - Scottish Nephrops trawlers



Figure 4.7.1 Nephrops, Moray Firth (FU 9), Long term landings and mean sizes. Note that the effort and LPUE from Scottish trawlers cover a shorter period 2000-2015.


Figure 4.7.2 Nephrops, Moray Firth (FU 9), Landings by quarter and sex from Scottish Nephrops trawlers.


Figure 4.7.3 Nephrops Moray Firth (FU 9) Length composition of catch of males (right) and females left from 2000 (bottom) to 2015 (top). Mean sizes of catch and landings are displayed vertically.


Figure4.7.4 Nephrops, Moray Firth (FU 9). TV survey distribution and relative density (2010-2015). Green and brown areas represent areas of suitable sediment for Nephrops. Density proportional to circle radius. Red crosses represent zero observations.


Figure 4.7.5 Nephrops, Moray Firth (FU 9), Time series of TV survey abundance estimates with 95\% confidence intervals, 1993-2015.


Figure 4.7.6 Nephrops, Moray Firth (FU 9), Quarterly sex ratio (by number) in catches.


Figure 4.8.1 Nephrops, Noup (FU 10), Long term landings, effort, LPUE and mean sizes (no females in samples in 2010 and no samples in 2015).


Figure 4.8.2 Nephrops, Noup (FU 10), Landings, effort (days) and LPUE (kg/day) split by TR1 and TR2 gears, data from year 2000.


Figure 4.8.3 Nephrops, Noup (FU 10). TV survey distribution and relative density (1994, 1999, 2006, $2007 \& 2014$ ). Green and brown areas represent areas of suitable sediment for Nephrops. Density proportional to circle radius. Red crosses represent zero observations.


Figure 4.8.4 Nephrops, Noup (FU 10), Time series of TV survey abundance estimates (absolute conversion factor $=1.35$, from Fladen), with $95 \%$ confidence intervals, 1994, 1999, 2006-2007 \& 2014.


Figure 4.9.1. Nephrops Norwegian Deep (FU 32). Catches and landings, Danish effort, Danish LPUE, and mean size in Danish discards and landings.



Figure 4.9.2. Nephrops Norwegian Deep (FU 32). Danish landings of Nephrops per ICES square. Dots represent hauls with Nephrops in at-sea-sampling program.

## Length frequencies for catch: Nephrops in FU32



Figure 4.9.3. Nephrops Norwegian Deep (FU 32). Size distribution in Danish catches.


Figure 4.9.4. Nephrops Norwegian Deep (FU 32). Size distribution of Danish and Norwegian catches. Vertical line indicates MLS ( 40 mm CL). Data from the Norwegian coast guard.


Figure 4.9.5. Nephrops Norwegian Deep (FU 32). Comparison of size distribution in catches (2006, $\mathbf{2 0 0 7}, 2012$ ) from Danish and Norwegian data sources. Vertical line indicates MLS ( 40 mm CL ).


Figure 4.9.6. Nephrops Norwegian Deep (FU 32). Positions of single trawl hauls with Nephrops in the catch from Norwegian bottom trawlers $\geq 15 \mathrm{~m}, 2011$-2015.


Figure 4.9.7. Nephrops Norwegian Deep (FU 32). Distribution of Nephrops in Norwegian shrimp survey, 2006-2016.


Figure 4.10.1 Nephrops in FU 33 (Off Horns Reef): Landings, effort and mean size.

## Length frequencies for catch:

Nephrops in FU33


Figure 4.10.2 Nephrops in FU 33 (Off Horn's Reef): Size distribution in Danish catches.


Figure 4.11.1. Nephrops, Devil's Hole (FU 34). British Geological Survey (BGS) map of sediment suitable for Nephrops in the northern North Sea. The Devil's Hole is located between 0 and 2 degrees east and 56 and 57.5 degrees north. Olive - muddy sand, lime green - sandy mud, dark green - mud.


Figure 4.11.2. Nephrops, Devil's Hole (FU 34). International landings (top) and Scottish landings split by TR1 and TR2 (bottom).

## LPUE



Figure 4.11.3. Nephrops, Devil's Hole (FU 34). Effort (days) and LPUE (kg/day) by Scottish trawlers split by TR1 and TR2 gears.


Figure 4.11.4. Nephrops, Devil's Hole (FU 34). UWTV survey distribution and relative density (20092015). Survey station locations generated from Vessel Monitoring System (VMS) data (WKNEPH, 2013). Density proportional to circle radius.


Figure 4.11.5. Nephrops, Devil's Hole (FU 34). Time series of UWTV survey density estimates with $95 \%$ confidence intervals, 2003, 2005, 2009-2015.


Figure 4.11.6. Nephrops, Devil's Hole (FU 34). Comparison of BGS sediment map with VMS data from Scottish trawlers (2007-2011) filtered for Nephrops landings $>30 \%$ of total, speeds of 0.5 - 3.8 knots and mesh size 70-99 mm.


Figure 4.11.7. Nephrops, Devil's Hole (FU 34). Union of 2007-2011 annual VMS polygons (from alpha convex hull) with VMS data filtered for Nephrops landings $>30 \%$ of total, speeds of 0.5 - 3.8 knots and mesh size 70-99 mm.

## 5 Norway Pout (Trisopterus esmarkii) in Subarea 4 and Division 3.a (North Sea, Skagerrak and Kattegat)

The advice, as well as the report, on Norway Pout will be released on the 11 November 2016.

## 6 Plaice (Pleuronectes platessa) in Division 7.d (Eastern English Channel)

This stock is in category 1. This year, the assessment of plaice in Division 7.d was made following methodological information described in the Stock Annex revised during ICES WKPLE 2015 and WGNSSK 2015.

### 6.1 General

### 6.1.1 Stock definition

A summary of available information can be found in the stock annex.

### 6.1.2 Ecosystem aspects

No new information on ecosystem aspects was presented at the working group in 2016. All available information on ecological aspects can be found in the Stock Annex.

### 6.1.3 Fisheries

Plaice is mainly caught in two offshore fisheries, i.e. the beam trawl sole fishery and the mixed demersal fishery using otter trawls. There is also a directed fishery during parts of the year by inshore trawlers and netters on the English and French coasts. All available information on the fisheries can be found in the Stock Annex.

### 6.1.4 ICES advices for previous years

2014 advice: Based on the ICES approach for data limited stocks, ICES advises that landings of plaice in Division 7.d plaice stock should be no more than 2811 tonnes. Assuming the same proportion of the Division 7.e and Subarea IV plaice stocks is taken in Division 7.d as during the last decade (2001-2012), this will correspond to total landings of plaice in Division 7.d of no more than 3469 tonnes.

2015 advice: ICES advises that when the MSY approach is applied, catches of the Division 7.d plaice stock in 2016 should be no more than 16923 tonnes. If discard rates do not change from the average (2012-2014), this implies landings of the Division 7.d plaice stock of no more than 10855 tonnes. Assuming the same proportion of the Division 7.e and Subarea IV plaice stocks is taken in Division 7.d as during 2003-2014, this will correspond to catches of plaice in Division 7.d in 2016 of no more than 19506 tonnes. If discard rates do not change from the average (2012-2014), this implies landings of plaice in Division 7.d of no more than 12512 tonnes.

### 6.1.5 Management

There are no explicit management objectives for this stock.
The TACs have been set to for the combined ICES Divisions 7.d \& 7.e.
The minimum landing size for plaice is 27 cm , which is not in accordance with the minimum mesh size of 80 mm , permitted for catching plaice by beam and otter trawling. Fixed nets are required to use $90-\mathrm{mm}$ mesh as an absolute minimum.

Demersal fisheries in the area are mixed fisheries, with many stocks exploited together in various combinations in the various fisheries. In these cases, management advice must consider both the state of individual stocks and their simultaneous exploitation in demersal fisheries. Stocks in the poorest condition, particularly those which suffer
from reduced reproductive capacity, become the overriding concern for the management of mixed fisheries, where these stocks are exploited either as a targeted species or as a bycatch.

### 6.2 Data available

### 6.2.1 Catch

Landings data as reported to ICES are shown in Figure 6.2.1.1 as well as in Table 6.2.1.1 together with the total landings estimated by the Working Group. The 2015 landings of 3727 t ( 2956 t attributed to the resident stock and 771 t removed from the first quarter as estimated to be resulting from catches coming from 7.e and IV to spawn) are in the catch level of the past 10 years (between 3500 and 4500 t ). Unlike previous years, France (45\%) and Belgium ( $43 \%$ ) contributed almost equally to the total 7.d landings in 2015, with UK contributing for $11 \%$.

Routine discard monitoring has recently begun following the introduction of the EU data collection regulations. Based on the sampling intensity (WKPLE 2015), a discards time series starting in 2006 has been included in the assessment.

Following the ICES WKFLAT 2010 and WKPLE 2015 conclusions, 65\% of the first quarter catches were removed. These 65\% were estimated during ICES WKFLAT 2010, based on published tagging results and some previous studies (e.g. Burt et al. 2006, Hunter et al. 2004, Kell et al. 2004) showing that $50 \%$ of the fish caught during the first quarter are fish coming from area IV to spawn. The same study also shown that $15 \%$ of the fish caught during the first quarter were fishes from area 7.e. Following the ICES WKPLE 2015 conclusions, only mature individuals are removed, both from landings and discards. Table 6.2.1.2 shows the Quarter 1 landings and discards and the corresponding removals. Removing this part of the catches allows for assessing the stock resident biomass. All the following figures will take into account this Quarter 1 removal.

### 6.2.2 Intercatch

UK, France, the Netherlands and Belgium have been providing landings data under the ICES InterCatch format since 2011, and InterCatch was used to produce the input data. Age distributions were provided by France, Belgium and England, accounting for $85 \%$ of the landings (Figure 6.2.2.1). Belgium has not always been able to provide landings data per quarter (for 2004, 2005, 2006, 2011, 2013, catch data were provided per semester or year), but they now provide it at least for quarter 1 on a separate excel spreadsheet. Allocations to calculate age structures for the remaining landings were done per quarter, using the groups below.

| UnSAMPLED FLEET* | SAMPLED FLEET** |
| :--- | :--- |
| All nets | All nets |
| All OTB, TBB and Seines | All OTB, TBB and Seines |
| Others (MIS and LLS) | All métiers |
| ${ }^{*}$ Unsampled fleet are those fleets for which no age structure is known. |  |
| ${ }^{* *}$ Sampled fleet are those fleets for which the age structure is known. |  |

Discards data have also been provided under the ICES InterCatch format by France, Belgium, and the UK since WKPLE (ICES, 2015). In 2015, $87 \%$ of landings had associated discards data imported to Intercatch. The discard volumes of the remaining strata have been raised using the grouping below (all quarters were pooled). As a result, the raised discards account for $14 \%$ of the total discards.

| UnSAMPLED FLEET* | SAMPLED FLEET** |
| :--- | :--- |
| TBB | TBB |
| GNS | GNS except 2 UK strata with high discards ratios |
| GTR | GTR |
| OTB | OTB |
| Seines (SDN and SSC) | Seines |
| Others (MIS and LLS) | All métiers |
| * Unsampled fleet are those fleets for which no discards data have been provided. |  |
| ${ }^{* *}$ Sampled fleet are those fleets for which the discards volumes are known. |  |

Age distributions were provided by France, Belgium and England, accounting for 79\% of the total discards (imported + raised).

### 6.2.3 Age compositions

Age compositions of the landings and of the discards are presented in Table 6.2.3.1 and Figure 6.2.3.1, and Table 6.2.3.3 and Figure 6.2.3.2 respectively.

Age distributions (exploitation pattern) may be quite different between quarters, as shown for 2015 in Figure 6.2.3.3, with recruits at age 0 and 1 starting to be caught and age 1 landed after summer.

Figure 6.2.3.4 presents the discards at age ratios (i.e. discards numbers/landings numbers) per age over the sampled period 2006-2015. From 2012, the ratio is higher for the ages 1 to 4 . The ratio for age 5 also increased to more than $20 \%$ in 2015.

### 6.2.4 Weight at age

Weights at age in the landings, in the discards and in the stock are presented in Table 6.2.3.2, 6.2.3.4 and 6.2.3.5 respectively and in Figure 6.2.3.1. Stock weights are assumed to be the Q2 landings weights. These weights at age do not show specific trends, apart from a general decrease in stock weights in 2013-2015.

### 6.2.5 Maturity and natural mortality

The maturity ogive used in the assessment is given in the table below.

| AGE | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Proportion of mature | 0 | 0.15 | 0.53 | 0.96 | 1 | 1 | 1 |

New age-specific natural mortality rates have been estimated from Peterson and Wroblewski's relationship during the 2015 WKPLE benchmark, as detailed in the Stock Annex.

| AGE | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Natural mortality | 0.3531 | 0.3132 | 0.292 | 0.2749 | 0.2594 | 0.2474 | 0.2329 |

### 6.2.6 Surveys

The survey series used in the assessment are the French Ground Fish Survey (FR GFS) and the UK beam trawl survey (UK BTS) (Figure 6.2.6.1 and Table 6.2.6.1). The International Young fish survey is also presented, although not used in the assessment. They are fully described in the stock annex.

New time series have been provided for both the FR GFS (Travers-Trolet et al, 2016; see Annex 08 under Working documents) and the UK BTS (Silva, 2016; see Annex 08 under Working documents).

The differences between the old and new time series are presented in Figure 6.2.6.2. The differences are very small for the UK BTS, while ages 1,2 and 6 of the FR GFS changed a bit more. In the case of age 1, the FR GFS indices is now closer to the UK BTS (Figure Figure 6.2.6.1). The consistencies of the two new surveys are presented in Figure 6.2.6.3. They are increased for ages 1 to 2 and 4 to 5 but slightly decreased for ages 2 to 3 and 3 to 4 of the FR GFS survey. They are increased for all ages but 1 to 2 of the UK BTS survey.

The effect of the changes in the survey indices on 2015 the assessment is presented in Figure 6.2.6.4. The use of the new indices leads to a slight change in the perception of the stock with a reduced recruitment, an increased fishing mortality, and a reduced SSB (around -15\% in 2014).

### 6.3 Assessment

The model used is the Art and Poos model (AAP, Aarts and Poos, 2009, for more details please refer to the Stock Annex).

|  | YeAr OF ASSESSMENT: |  |
| :--- | :--- | :--- |
| Assessment model: | 2016 |  |
| Assessment software | AAP |  |
| Fleets: | FLR/ADMB |  |
| UK Beam Trawl Survey |  |  |
| Year range | Age range |  |
| FR Ground Fish Survey |  |  |
| Year range | Age range | $1-6$ |
| Catch/Landings | 1988 onwards |  |
| Age range: | $1-6$ |  |
| Landings data: | 1988 onwards |  |
| Discards data | $1-7+$ |  |
| Model settings | $1980-2015$ |  |
| Fbar: | $2006-2015$ |  |
| Age from which F is constant (qplat.Fmatrix) | $3-6$ |  |
| Dimension of the F matrix (Fage.knots) | 6 |  |
| Ftime.knots | 4 |  |
| Wtime.knots | 14 |  |
| Age from which q is constant (qplat.surveys) | 5 |  |

### 6.3.1 Results

The landings and discards estimated by the model are presented in Figure 6.3.1.1 and the residuals in Tables 6.3.1.1 and 6.3.1.2. As last year, given the observed trend in the discard at age ratio (see section 6.2.3), the actual discard at age ratio (rather than the average one, i.e. the black line on the left panel of Figure 6.2.1.3) are used in the assessment to estimate the discards for the last 3 years (2012 to 2015).
The survey residuals are shown in Figure 6.3.1.2 and Table 6.3.1.3for the two surveys. There are opposite trends in the residuals of the UK BTS and French GFS (the two surveys covering the entire geographical area of the stock) appearing in the late 2000s, particularly for ages 1 and 2 .
The final outputs are given in Table 6.3.1.4 (fishing mortalities) and Table 6.3.1.5 (stock numbers). A summary of the assessment results is given in Table 6.3.1.6 and trends in fishing mortality, recruitment, spawning stock and total catches are shown in Figure 6.3.1.3. Retrospective patterns for the final run are shown in Figure 6.3.1.4. The model tends to underestimate the recruitment.

The 1986 year class dominated the history of this stock until the late 2000s (Figure 6.3.1.5 and 6.3.1.3). A second peak occurred with the 1997 year class, although estimated to be at $75 \%$ of the 1986 year class. The ephemeral peak of SSB in 1999 has been followed by years of stability at a low level. This low SSB situation was confirmed by the fisher's perception and assessed by a survey in France in 2006. The SSB has now been increasing for the last 5 years. From 2006 onwards, a series of high recruitments occurred, reaching a maximum in 2010, which caused to biomass to increase until now (Figure 6.3.1.5).

### 6.4 Biological reference points

FMSY was estimated last year using the procedure advised during WKMSYREF3 2014 (WGNSSK, 2015). Three stock-recruitment relationships were assessed which led to the selection of the hockey-stick and the Beverton and Holt models. Then, FMSY was determined using the eqsim method from the R library MSY.

### 6.4.1 Calculation of additional reference points

This year, $\mathrm{F}_{\mathrm{lim}}$ (Figure 6.4.1.1) and $\mathrm{F}_{\mathrm{pa}}$ were calculated according to the recommendations from ACOM (ICES, 2016).

| References points | Value | Notes |
| :--- | :--- | :--- |
| $\mathrm{F}_{\text {lim }}$ | 0.50 | The fishing mortality (F) <br> that in equilibrium will <br> maintain the stock above <br> Blim with a 50\% probabil- <br> ity. |
|  |  | $\mathrm{F}_{\text {lim }} / 1.4$ |
| $\mathrm{~F}_{\mathrm{pa}}$ default | $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} \times \exp (-1.645 \times \sigma)$ <br> where $\sigma$ is the standard <br> error of $\ln (\mathrm{F})$ in the final <br> assessment year. |  |

### 6.5 Short-term forecasts

Weight-at-age in the stock and in the catch were taken to be the average over the last 3 years. The exploitation pattern, as well as the discards/landings numbers ratio, were taken to be the mean value of the last three years. Population numbers at age 2 and older in 2014 are AAP survivors estimates.

### 6.5.1 Recruitment estimates

Considering the retrospective patterns observed, the recruitment is assumed to be poorly estimated.

For 2016 and the previsions (2017 and 2018), the recruitment was calculated as the geometric mean recruitment over the period 2010-2013 (taking into account the higher recruitment in the recent years

### 6.5.2 Calculation of the 7.d resident stock

F for the intermediate year is set such as landings equal the TAC for that year. However, TAC is combined for area 7.d and 7.e. The long term proportion of catches taken in area 7.d over the total catches is used to compute a TAC 7.d (Figure 6.5.2.1).

As catch numbers and AAP survivors in 2015 are computed from the resident population (catches made on fishes from area 7.e and IV are removed), the TAC in 7.d and resulting F in intermediate year were also modified to take into account this first quarter removal. The first quarter removal was also estimated as the long term average, and TAC reduced by this average (Figure 6.5.2.1).

### 6.5.3 Management options tested

### 6.5.3.1 Calculation of STF

Potential TACs for 2017 were calculated using Fmsy and the newly estimated Fpa. Alternative options were also tested. Results are presented in Table 6.5.3.1.1

Following the MSY approach would lead to a TAC in 2017 for the resident stock of 12805 t , corresponding to an estimated wanted catch of 7550 t , i.e. an increase compared to the 2015 landings.

### 6.6 Quality of the assessment

The sampling for plaice in 7.d are considered to be at a reasonable level.
The quality of the assessment is considered to have improved in 2015 following the change of assessment model and the inclusion of discards. Some concerns however were expressed during the group about the change of natural mortality rate values which leads to a significant change in the perception of this stock. The assessment was therefore externally reviewed, and the new mortality rates maintained.

A fishery on the spawners takes place during the first quarter of the year, yielding an age distribution different from the rest of the year. It is unknown whether there is major inter-annual variability in the immigration from the North Sea to these spawning grounds, which could distort any catch-based analysis. Any migration events taking place in the first quarter cannot be represented in the surveys in the second semester.
Both landings-at-age and tuning fleets information are highly dependent on the accuracy of the spatial declaration of the fishing activity as an important component of the fisheries operates on the borderline to ICES subdivision 4.c.

### 6.7 Status of the stock

Results of the assessment indicate that F is stable at a low level, while SSB has been increasing in recent years.

### 6.8 Management considerations

The stock identity of plaice in the Channel is unclear and may raise some issues.
The TAC is combined for Divisions 7.d and 7.e. Plaice in 7.e is considered at risk of being harvested unsustainably ( F above $\mathrm{F}_{\text {msy, }}$, although the ssb is estimated to be increasing as in 7.d).

The plaice stock in 7.d is mostly harvested in a mixed fishery with sole in 7.d.
Due to the minimum mesh size $(80 \mathrm{~mm})$ in the mixed beam and otter trawl fisheries, a large number of undersized plaice are discarded. The 80 mm mesh size is not matched to the minimum landing size of plaice $(27 \mathrm{~cm})$. Measures taken specifically to control sole fisheries will impact the plaice fisheries.

### 6.9 Frequency of assessment

In this section, the criteria defining whether a stock is candidate for less frequent assessments (ACOM, 2014) are assessed for ple-eche.

| Stocks are considered candidates for BIENNIAL ASSESSMENT IF: | PLE-ECHE |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| The advice for the stock has been 0catch or equivalent for the latest three advice years | No |  |  |  |  |  |
| The following criteria are fulfilled simultaneously: |  |  |  |  |  |  |
| Life span (i.e. maximum normal age) of the species is larger than 5 years | Yes |  |  |  |  |  |
| The stock status in relation to the reference points is according to the MSY criteria F (latest assessment year) $<=1.1 \times$ Fmsy OR if $\mathrm{F}_{\text {msy }}$ range has been defined: F (latest assessment year) is $<=$ Fupper (upper bound in F range) AND SSB (start of intermediate year) $>=$ MSY $_{\text {trigger }}$ | Yes$\mathrm{F}(2015)$ <br> 0.12 |  | $1.1 \times \mathrm{F}$ $1.1 \times 0.2$ | y $=0.275$ |  |  |
| The average contribution to the catch in numbers of the recruiting year class in latest 5 years is less than $25 \%$ of the total catch in numbers. Should be calculated as the average over the latest five years of the catch in numbers of first age divided by the total catch in number by year. | Yes <br> Year <br> \% age 1 <br> caught / total <br> catch | 2011 | 2012 $5 \%$ | 2013 <br> $7 \%$ | 2014 $16 \%$ | 2015 $15 \%$ |
| The retrospective pattern, based on a seven years peel of Mohn's Rho index, shows that F is consistently underestimated by less than $20 \%$ <br> The formula to be used in the calculations is: <br> $\rho=\frac{1}{7} \sum_{u=Y-7}^{Y-1}\left(1-\frac{F_{u, u}}{F_{u, Y}}\right)$. The result should be $<0.20$, where $\boldsymbol{F}_{u, u}$ is F in year u estimated from an assessment that ends in year u , and $\boldsymbol{F}_{u, Y}$ is the F in year u estimated from the most recent assessment (which ends in year Y ) | Results from $W$  <br> Yes, $\varrho=0.11$  <br> 2007 2008 <br> 0.35 0.16 <br> Results from W Yes, $\mathrm{Q}=0.04$ <br> 20112012 <br> $0.24 \quad 0.12$ <br> The model cann 2011 due to a sm trends in discar | NSSK <br> 2009 <br> -0.03 <br> NSSK <br> 2013 <br> -0.03 <br> be us <br> ll chan <br> at ag | 2015 <br> 2010 <br> 0.11 <br> 2016 <br> 2014 <br> -0.16 <br> d to ru ge in th ratio. | 2011 <br> 0.17 <br> simu <br> code | $\begin{aligned} & 2012 \\ & -0.03 \end{aligned}$ <br> tions p accoun | $\begin{aligned} & 2013 \\ & 0.03 \end{aligned}$ <br> ior to t for |

### 6.10 Sources

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Table 6.2.1.1 - Plaice in 7d. Nominal landings (tonnes) as officially reported to ICES, 1976-2014.

| $\underset{\underset{\sim}{\underset{\sim}{4}}}{\substack{\sim}}$ | $\underset{\sim}{山}$ | $\underset{\sim}{4}$ |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 147 | 1439 | 376 | 1(1) | 1963 | - | 1963 |  | 640 |  |
| 1977 | 149 | 1714 | 302 | 81(2) | 2246 | - | 2246 |  | 702 |  |
| 1978 | 161 | 1810 | 349 | 156(2) | 2476 | - | 2476 |  | 784 |  |
| 1979 | 217 | 2094 | 278 | 28(2) | 2617 | - | 2617 |  | 977 |  |
| 1980 | 435 | 2905 | 304 | 112(2) | 3756 | -1106 | 2650 |  | 1215 |  |
| 1981 | 815 | 3431 | 489 | - | 4735 | 34 | 4769 |  | 1746 |  |
| 1982 | 738 | 3504 | 541 | 22 | 4805 | 60 | 4865 |  | 1938 |  |
| 1983 | 1013 | 3119 | 548 | - | 4680 | 363 | 5043 |  | 1754 |  |
| 1984 | 947 | 2844 | 640 | - | 4431 | 730 | 5161 |  | 1813 |  |
| 1985 | 1148 | 3943 | 866 | - | 5957 | 65 | 6022 |  | 1751 |  |
| 1986 | 1158 | 3288 | 828 | 488 (2) | 5762 | 1072 | 6834 |  | 2161 |  |
| 1987 | 1807 | 4768 | 1292 | - | 7867 | 499 | 8366 |  | 2388 | 8300 |
| 1988 | 2165 | 5688 (2) | 1250 | - | 9103 | 1317 | 10420 |  | 2994 | 9960 |
| 1989 | 2019 | 3265 (1) | 1383 | - | 6667 | 2091 | 8758 |  | 2808 | 11700 |
| 1990 | 2149 | 4170 (1) | 1479 | - | 7798 | 1249 | 9047 |  | 3058 | 10700 |
| 1991 | 2265 | 3606 (1) | 1566 | - | 7437 | 376 | 7813 |  | 2250 | 10700 |
| 1992 | 1560 | 3099 | 1553 | 20 | 6232 | 105 | 6337 |  | 1950 | 9600 |
| 1993 | 877 | 2792 | 1075 | 27 | 4771 | 560 | 5331 |  | 1691 | 8500 |
| 1994 | 1418 | 3199 | 993 | 23 | 5633 | 488 | 6121 |  | 1471 | 9100 |
| 1995 | 1157 | 2598 (2) | 796 | 18 | 4569 | 561 | 5130 |  | 1295 | 8000 |
| 1996 | 1112 | 2630 (2) | 856 | + | 4598 | 795 | 5393 |  | 1321 | 7530 |
| 1997 | 1161 | 3077 | 1078 | + | 5316 | 991 | 6307 |  | 1654 | 7090 |
| 1998 | 854 | 3603 (2) | 700 | + | 5157 | 605 | 5762 |  | 1430 | 5700 |
| 1999 | 1306 | 3388 (1) | 743 | + | 5437 | 889 | 6326 |  | 1616 | 7400 |
| 2000 | 1298 | 3183 | 752 | + | 5233 | 782 | 6015 |  | 1678 | 6500 |
| 2001 | 1346 | 2962 | 655 | + | 4963 | 303 | 5266 |  | 1379 | 6000 |
| 2002 | 1204 | 3454 | 841 |  | 5499 | 278 | 5777 |  | 1608 | 6700 |
| 2003 | 998 | 2893 | 756 | 3 | 4650 | -564 | 4086 |  | 1478 | 6000 |
| 2004 | 954 | 2766 | 582 | 10 | 4312 | 438 | 4750 |  | 1402 | 6060 |
| 2005 | 832 | 2432 | 421 | 21 | 3706 | 285 | 3991 |  | 1370 | 5150 |
| 2006 | 1024 | 1935 | 550 | 16 | 3525 | 121 | 3646 | 749 | 1466 | 5080 |
| 2007 | 1355 | 2017 | 463 | 10 | 3845 | 156 | 4001 | 1252 | 1184 | 5050 |
| 2008 | 1386 | 1740 | 471 | 12 | 3609 | 255 | 3864 | 936 | 1144 | 4646 |
| 2009 | 1002 | 1892 | 612 | 16 | 3522 | 38 | 3560 | 1528 | 1043 | 4274 |
| 2010 | 1123 | 2190 | 517 | 62 | 3892 | 519 | 4411 | 2511 | 2240 | 4665 |
| 2011 | 1067 | 1994 | 472 | 56 | 3589 | 60 | 3649 | 2025 | 1192 | 4665 |
| 2012 | 1045 | 1962 | 542 | 63 | 3612 | 111 | 3723 | 3336 | 1339 | 5092 |
| 2013 | 1295 | 2159 | 641 | 87 | 4182 | -55 | 4127 | 2955 | 1526 | 6400 |
| 2014 | 1389 | 2229 | 633 | 76 | 4327 | -7 | 4320 | 3886 | 1339 | 5322 |
| 2015 | 1605 | 1664 | 390 | 53 | 3712 | 15 | 3727 | 2821 |  | 6223 |

Estimated by the working group from combined Division 7d+e
Includes Division 7e
As provided to ICES through InterCatch
Raised with InterCatch from BE, UK and FR estimated discards data.
TAC's for Divisions 7 d,e

Table 6.2.1.2 - Plaice in 7.d. Nominal landings, estimated discards, and Quarter 1 removals

| Year | Total LANDINGS | $\begin{gathered} \text { Q1 } \\ \text { Remov. } \end{gathered}$ | LANDINGS AS used by WG (1) | Estim. DISCARDS | DISCARDS Q1 REMOV. | DISCARDS AS USED By WG (1) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 2650 | 427 | 2223 |  |  |  |
| 1981 | 4769 | 760 | 4009 |  |  |  |
| 1982 | 4865 | 825 | 4040 |  |  |  |
| 1983 | 5043 | 950 | 4093 |  |  |  |
| 1984 | 5161 | 912 | 4249 |  |  |  |
| 1985 | 6022 | 1022 | 5000 |  |  |  |
| 1986 | 6834 | 1161 | 5673 |  |  |  |
| 1987 | 8366 | 1360 | 7006 |  |  |  |
| 1988 | 10420 | 1635 | 8785 |  |  |  |
| 1989 | 8758 | 1665 | 7093 |  |  |  |
| 1990 | 9047 | 1698 | 7349 |  |  |  |
| 1991 | 7813 | 1451 | 6362 |  |  |  |
| 1992 | 6337 | 1118 | 5220 |  |  |  |
| 1993 | 5331 | 852 | 4479 |  |  |  |
| 1994 | 6121 | 1074 | 5047 |  |  |  |
| 1995 | 5130 | 934 | 4196 |  |  |  |
| 1996 | 5393 | 963 | 4430 |  |  |  |
| 1997 | 6307 | 1127 | 5180 |  |  |  |
| 1998 | 5762 | 931 | 4832 |  |  |  |
| 1999 | 6326 | 1058 | 5268 |  |  |  |
| 2000 | 6015 | 1494 | 4522 |  |  |  |
| 2001 | 5266 | 886 | 4380 |  |  |  |
| 2002 | 5777 | 931 | 4846 |  |  |  |
| 2003 | 4086 | 476 | 3610 |  |  |  |
| 2004 | 4750 | 544 | 4206 |  |  |  |
| 2005 | 3991 | 506 | 3485 |  |  |  |
| 2006 | 3646 | 421 | 3225 | 749 | 21 | 727 |
| 2007 | 4001 | 620 | 3381 | 1252 | 32 | 1220 |
| 2008 | 3864 | 586 | 3278 | 936 | 48 | 888 |
| 2009 | 3560 | 436 | 3124 | 1528 | 56 | 1473 |
| 2010 | 4411 | 501 | 3910 | 2511 | 99 | 2412 |
| 2011 | 3649 | 358 | 3291 | 2025 | 99 | 1926 |
| 2012 | 3723 | 544 | 3179 | 3336 | 293 | 3043 |
| 2013 | 4127 | 523 | 3604 | 2955 | 260 | 2696 |
| 2014 | 4320 | 645 | 3675 | 3886 | 561 | 3325 |
| 2015 | 3727 | 771 | 2956 | 2821 | 453 | 2368 |

1. takes into account the removal of $\mathbf{6 5 \%}$ of the Quarter 1 landings or discards.

Table 6.2.3.1. Plaice in 7.d. Landings in numbers (thousands) as used in the assessment, taking into account the first quarter removal.

| AGE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1980 | 53 | 2598 | 1253 | 370 | 324 | 50 | 133 |
| 1981 | 16 | 2403 | 5866 | 1643 | 192 | 106 | 238 |
| 1982 | 265 | 1369 | 5964 | 2262 | 505 | 138 | 179 |
| 1983 | 92 | 2977 | 2761 | 4048 | 617 | 151 | 214 |
| 1984 | 350 | 1838 | 6310 | 1928 | 1242 | 356 | 312 |
| 1985 | 142 | 5614 | 5347 | 3346 | 274 | 409 | 300 |
| 1986 | 679 | 4799 | 6072 | 2510 | 965 | 375 | 247 |
| 1987 | 25 | 8350 | 6481 | 2379 | 833 | 287 | 512 |
| 1988 | 16 | 4923 | 16239 | 3357 | 741 | 362 | 561 |
| 1989 | 826 | 3574 | 6238 | 6477 | 1770 | 392 | 497 |
| 1990 | 1632 | 2581 | 7550 | 4099 | 2386 | 535 | 572 |
| 1991 | 1542 | 5758 | 4700 | 3099 | 1614 | 1123 | 429 |
| 1992 | 1665 | 6085 | 3841 | 1183 | 786 | 697 | 745 |
| 1993 | 740 | 7473 | 3295 | 863 | 359 | 313 | 581 |
| 1994 | 1242 | 3570 | 6015 | 2131 | 563 | 280 | 781 |
| 1995 | 2592 | 4264 | 2532 | 2006 | 611 | 152 | 591 |
| 1996 | 1119 | 4762 | 3113 | 1060 | 951 | 326 | 585 |
| 1997 | 550 | 4168 | 6184 | 2382 | 724 | 506 | 722 |
| 1998 | 464 | 4323 | 7467 | 2335 | 360 | 94 | 289 |
| 1999 | 741 | 1737 | 10493 | 4583 | 696 | 121 | 223 |
| 2000 | 1383 | 6177 | 3432 | 3992 | 752 | 150 | 142 |
| 2001 | 2682 | 4070 | 3589 | 1385 | 1253 | 203 | 145 |
| 2002 | 902 | 6876 | 4553 | 1390 | 1144 | 603 | 288 |
| 2003 | 0 | 3597 | 2103 | 1380 | 350 | 356 | 758 |
| 2004 | 922 | 2718 | 4573 | 760 | 400 | 219 | 527 |
| 2005 | 86 | 2602 | 2153 | 1975 | 449 | 245 | 508 |
| 2006 | 191 | 2801 | 3081 | 1626 | 987 | 166 | 379 |
| 2007 | 529 | 2986 | 2379 | 1237 | 534 | 395 | 274 |
| 2008 | 293 | 3844 | 2512 | 1125 | 584 | 218 | 258 |
| 2009 | 491 | 2975 | 3112 | 848 | 402 | 242 | 240 |
| 2010 | 530 | 4238 | 3367 | 1465 | 392 | 278 | 287 |
| 2011 | 93 | 4436 | 3557 | 964 | 316 | 59 | 119 |
| 2012 | 18 | 1266 | 3780 | 1845 | 524 | 195 | 171 |
| 2013 | 9 | 756 | 3666 | 3294 | 1158 | 247 | 156 |
| 2014 | 76 | 759 | 2015 | 3731 | 1848 | 468 | 202 |
| 2015 | 3 | 600 | 1523 | 1483 | 1933 | 940 | 642 |

Table 6.2.3.2. Plaice in 7.d. Weights in the landings.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.309 | 0.312 | 0.499 | 0.627 | 0.787 | 1.138 | 1.413 |
| 1981 | 0.224 | 0.280 | 0.349 | 0.434 | 0.666 | 0.814 | 1.000 |
| 1982 | 0.226 | 0.250 | 0.325 | 0.397 | 0.589 | 0.732 | 1.134 |
| 1983 | 0.239 | 0.267 | 0.314 | 0.378 | 0.488 | 0.740 | 1.112 |
| 1984 | 0.201 | 0.255 | 0.290 | 0.347 | 0.438 | 0.595 | 0.812 |
| 1985 | 0.235 | 0.257 | 0.278 | 0.395 | 0.464 | 0.527 | 0.798 |
| 1986 | 0.226 | 0.306 | 0.331 | 0.406 | 0.546 | 0.486 | 0.806 |
| 1987 | 0.245 | 0.275 | 0.351 | 0.465 | 0.563 | 0.764 | 0.947 |
| 1988 | 0.269 | 0.247 | 0.296 | 0.398 | 0.517 | 0.606 | 0.892 |
| 1989 | 0.187 | 0.250 | 0.299 | 0.345 | 0.441 | 0.604 | 1.008 |
| 1990 | 0.197 | 0.251 | 0.319 | 0.370 | 0.473 | 0.597 | 1.030 |
| 1991 | 0.217 | 0.267 | 0.299 | 0.375 | 0.437 | 0.535 | 0.978 |
| 1992 | 0.179 | 0.273 | 0.347 | 0.422 | 0.501 | 0.576 | 0.783 |
| 1993 | 0.219 | 0.270 | 0.334 | 0.429 | 0.504 | 0.587 | 0.860 |
| 1994 | 0.241 | 0.268 | 0.286 | 0.354 | 0.463 | 0.572 | 0.968 |
| 1995 | 0.214 | 0.266 | 0.307 | 0.382 | 0.476 | 0.675 | 0.927 |
| 1996 | 0.225 | 0.306 | 0.296 | 0.404 | 0.485 | 0.656 | 1.102 |
| 1997 | 0.194 | 0.246 | 0.291 | 0.324 | 0.431 | 0.562 | 0.990 |
| 1998 | 0.163 | 0.249 | 0.273 | 0.389 | 0.513 | 0.779 | 1.140 |
| 1999 | 0.196 | 0.244 | 0.234 | 0.305 | 0.460 | 0.749 | 1.093 |
| 2000 | 0.208 | 0.246 | 0.262 | 0.284 | 0.376 | 0.578 | 0.914 |
| 2001 | 0.225 | 0.263 | 0.317 | 0.387 | 0.467 | 0.671 | 1.093 |
| 2002 | 0.248 | 0.250 | 0.301 | 0.367 | 0.427 | 0.548 | 0.825 |
| 2003 | NA | 0.282 | 0.371 | 0.478 | 0.633 | 0.644 | 0.859 |
| 2004 | 0.246 | 0.298 | 0.400 | 0.499 | 0.690 | 0.788 | 0.997 |
| 2005 | 0.285 | 0.312 | 0.345 | 0.444 | 0.557 | 0.653 | 1.088 |
| 2006 | 0.259 | 0.277 | 0.304 | 0.361 | 0.444 | 0.552 | 0.843 |
| 2007 | 0.173 | 0.303 | 0.379 | 0.454 | 0.520 | 0.583 | 0.913 |
| 2008 | 0.231 | 0.282 | 0.338 | 0.418 | 0.529 | 0.623 | 0.939 |
| 2009 | 0.234 | 0.283 | 0.341 | 0.486 | 0.513 | 0.643 | 1.046 |
| 2010 | 0.225 | 0.296 | 0.348 | 0.438 | 0.497 | 0.639 | 0.822 |
| 2011 | 0.156 | 0.259 | 0.347 | 0.478 | 0.625 | 0.805 | 1.056 |
| 2012 | 0.198 | 0.287 | 0.346 | 0.437 | 0.540 | 0.691 | 1.026 |
| 2013 | 0.140 | 0.254 | 0.309 | 0.381 | 0.480 | 0.712 | 1.038 |
| 2014 | 0.169 | 0.249 | 0.283 | 0.357 | 0.492 | 0.674 | 0.945 |
| 2015 | 0.122 | 0.220 | 0.295 | 0.336 | 0.401 | 0.522 | 0.817 |

Table 6.2.3.3. Plaice in 7.d. Discards in numbers (thousands) as used in the assessment, taking into account the first quarter removal.

| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 553 | 2541 | 1826 | 70 | 10 | 1 | 0 |
| 2007 | 1227 | 5531 | 1776 | 278 | 0 | 2 | 0 |
| 2008 | 2368 | 2893 | 631 | 163 | 38 | 8 | 1 |
| 2009 | 2032 | 5679 | 1988 | 114 | 17 | 26 | 3 |
| 2010 | 2023 | 11797 | 3243 | 336 | 28 | 3 | 2 |
| 2011 | 2480 | 8872 | 1559 | 155 | 14 | 19 | 1 |
| 2012 | 1423 | 10296 | 7943 | 1235 | 52 | 0 | 0 |
| 2013 | 2040 | 5395 | 9367 | 1818 | 89 | 9 | 1 |
| 2014 | 4380 | 6222 | 8481 | 3445 | 493 | 79 | 10 |
| 2015 | 4420 | 8316 | 4958 | 1478 | 761 | 276 | 40 |

Table 6.2.3.4. Plaice in 7.d. Weights in the discards.

| YEAR | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2006 | 0.100 | 0.138 | 0.165 | 0.205 | 0.258 | 0.563 | NA |
| 2007 | 0.103 | 0.139 | 0.157 | 0.163 | 0.284 | 0.214 | NA |
| 2008 | 0.118 | 0.153 | 0.189 | 0.222 | 0.219 | 0.384 | NA |
| 2009 | 0.125 | 0.137 | 0.169 | 0.449 | 0.729 | 1.298 | 0.267 |
| 2010 | 0.103 | 0.135 | 0.167 | 0.180 | 0.237 | 0.381 | 0.369 |
| 2011 | 0.096 | 0.154 | 0.173 | 0.215 | 0.214 | 0.227 | 1.348 |
| 2012 | 0.092 | 0.129 | 0.165 | 0.192 | 0.211 | 0.602 | NA |
| 2013 | 0.081 | 0.125 | 0.151 | 0.184 | 0.243 | 0.453 | 0.411 |
| 2014 | 0.092 | 0.126 | 0.139 | 0.236 | 0.252 | 0.307 | 0.392 |
| 2015 | 0.039 | 0.105 | 0.155 | 0.173 | 0.219 | 0.273 | 0.618 |

Table 6.2.3.5. Plaice in 7.d. Weights in the stock.

| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.171 | 0.332 | 0.482 | 0.622 | 0.751 | 0.870 | 1.197 |
| 1981 | 0.110 | 0.216 | 0.317 | 0.414 | 0.506 | 0.594 | 0.924 |
| 1982 | 0.105 | 0.208 | 0.308 | 0.406 | 0.502 | 0.596 | 0.869 |
| 1983 | 0.097 | 0.192 | 0.286 | 0.379 | 0.470 | 0.560 | 0.854 |
| 1984 | 0.082 | 0.164 | 0.248 | 0.333 | 0.420 | 0.507 | 0.738 |
| 1985 | 0.084 | 0.171 | 0.259 | 0.348 | 0.440 | 0.533 | 0.778 |
| 1986 | 0.101 | 0.205 | 0.311 | 0.420 | 0.532 | 0.646 | 0.850 |
| 1987 | 0.122 | 0.242 | 0.361 | 0.479 | 0.596 | 0.712 | 0.929 |
| 1988 | 0.084 | 0.168 | 0.254 | 0.340 | 0.427 | 0.514 | 0.715 |
| 1989 | 0.079 | 0.162 | 0.250 | 0.342 | 0.439 | 0.541 | 0.855 |
| 1990 | 0.085 | 0.230 | 0.322 | 0.346 | 0.465 | 0.549 | 1.118 |
| 1991 | 0.143 | 0.219 | 0.275 | 0.335 | 0.375 | 0.472 | 0.958 |
| 1992 | 0.088 | 0.241 | 0.336 | 0.421 | 0.477 | 0.521 | 0.725 |
| 1993 | 0.108 | 0.258 | 0.296 | 0.379 | 0.493 | 0.539 | 0.727 |
| 1994 | 0.165 | 0.198 | 0.276 | 0.331 | 0.383 | 0.493 | 0.866 |
| 1995 | 0.124 | 0.257 | 0.286 | 0.354 | 0.442 | 0.707 | 0.855 |
| 1996 | 0.178 | 0.229 | 0.263 | 0.347 | 0.354 | 0.474 | 0.934 |
| 1997 | 0.059 | 0.202 | 0.256 | 0.266 | 0.417 | 0.530 | 0.902 |
| 1998 | 0.072 | 0.203 | 0.273 | 0.361 | 0.530 | 0.670 | 0.873 |
| 1999 | 0.072 | 0.172 | 0.213 | 0.351 | 0.429 | 0.644 | 0.904 |
| 2000 | 0.068 | 0.184 | 0.204 | 0.246 | 0.355 | 0.554 | 0.928 |
| 2001 | 0.093 | 0.206 | 0.274 | 0.338 | 0.404 | 0.624 | 1.104 |
| 2002 | 0.102 | 0.206 | 0.281 | 0.379 | 0.467 | 0.558 | 0.809 |
| 2003 | NA | 0.306 | 0.403 | 0.528 | 0.673 | 0.592 | 0.961 |
| 2004 | 0.280 | 0.366 | 0.508 | 0.571 | 0.701 | 0.788 | 0.861 |
| 2005 | 0.174 | 0.299 | 0.377 | 0.489 | 0.672 | 0.683 | 1.010 |
| 2006 | 0.220 | 0.270 | 0.343 | 0.419 | 0.506 | 0.637 | 0.938 |
| 2007 | 0.063 | 0.247 | 0.391 | 0.543 | 0.579 | 0.656 | 0.825 |
| 2008 | 0.121 | 0.245 | 0.301 | 0.368 | 0.448 | 0.462 | 1.005 |
| 2009 | NA | 0.268 | 0.358 | 0.487 | 0.476 | 0.719 | 1.036 |
| 2010 | NA | 0.280 | 0.354 | 0.415 | 0.455 | 0.561 | 0.719 |
| 2011 | 0.189 | 0.238 | 0.402 | 0.535 | 0.737 | 0.791 | 0.908 |
| 2012 | NA | 0.253 | 0.298 | 0.424 | 0.517 | 0.629 | 0.938 |
| 2013 | 0.174 | 0.252 | 0.277 | 0.479 | 0.454 | 0.886 | 0.995 |
| 2014 | 0.157 | 0.256 | 0.243 | 0.381 | 0.518 | 0.756 | 1.042 |
| 2015 | 0.154 | 0.253 | 0.256 | 0.287 | 0.363 | 0.436 | 0.782 |

Table 6.2.6.1. Plaice in 7.d. Tuning fleets

| UK BTS |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19892015 |  |  |  |  |  |  |
| 110.50 .75 |  |  |  |  |  |  |
| 16 |  |  |  |  |  |  |
| 1 | 3.8 | 15.8 | 28.9 | 31.7 | 4.0 | 1.7 |
| 1 | 9.2 | 9.4 | 11.1 | 11.7 | 12.6 | 1.5 |
| 1 | 16.8 | 14.5 | 11.5 | 8.7 | 8.6 | 4.6 |
| 1 | 22.4 | 21.3 | 6.6 | 6.6 | 7.2 | 5.4 |
| 1 | 4.6 | 20.2 | 8.0 | 2.8 | 2.9 | 2.4 |
| 1 | 9.4 | 8.5 | 10.1 | 6.0 | 2.0 | 0.6 |
| 1 | 14.5 | 6.2 | 3.8 | 5.7 | 2.2 | 0.8 |
| 1 | 22.1 | 17.3 | 1.7 | 1.0 | 2.0 | 1.3 |
| 1 | 48.2 | 28.6 | 11.0 | 1.3 | 1.6 | 0.5 |
| 1 | 30.6 | 37.9 | 12.1 | 5.0 | 0.6 | 0.6 |
| 1 | 12.8 | 10.7 | 28.8 | 4.6 | 1.6 | 0.3 |
| 1 | 19.5 | 30.2 | 18.8 | 20.5 | 5.0 | 1.3 |
| 1 | 27.9 | 20.3 | 14.1 | 9.8 | 14.8 | 2.7 |
| 1 | 37.9 | 25.9 | 12.5 | 5.5 | 2.6 | 5.3 |
| 1 | 10.6 | 39.7 | 9.8 | 4.4 | 2.3 | 1.1 |
| 1 | 52.9 | 22.5 | 20.7 | 4.8 | 1.2 | 0.3 |
| 1 | 15.6 | 36.2 | 12.8 | 10.0 | 3.2 | 1.1 |
| 1 | 30.1 | 28.9 | 16.8 | 5.9 | 4.3 | 1.3 |
| 1 | 53.1 | 28.9 | 12.2 | 6.2 | 3.2 | 2.9 |
| 1 | 39.6 | 40.6 | 10.5 | 4.3 | 3.8 | 1.8 |
| 1 | 77.7 | 39.5 | 20.9 | 5.9 | 3.2 | 2.3 |
| 1 | 64.2 | 64.7 | 17.7 | 9.2 | 3.1 | 1.7 |
| 1 | 115.1 | 112.2 | 39.6 | 10.3 | 7.0 | 2.9 |
| 1 | 24.7 | 81.1 | 56.0 | 18.7 | 4.2 | 3.3 |
| 1 | 32.3 | 61.0 | 88.2 | 45.0 | 10.2 | 3.4 |
| 1 | 145.3 | 156.5 | 50.7 | 62.1 | 26.8 | 9.0 |
| 1 | 38 | 178.7 | 63.2 | 30.2 | 33.4 | 15.7 |

Table 6.2.6.1.(cont.) Plaice in 7.d. Tuning fleets

| FR GFS |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19932015 |  |  |  |  |  |  |
| 110.751 |  |  |  |  |  |  |
| 16 |  |  |  |  |  |  |
| 1 | 232.04 | 867.4 | 345 | 125.8 | 32 | 8.66 |
| 1 | 468.69 | 347.5 | 148 | 67.6 | 26.2 | 11.65 |
| 1 | 30.31 | 336.5 | 364 | 142.1 | 101.1 | 27.19 |
| 1 | 772.65 | 243.8 | 181 | 26.6 | 12.9 | 15.07 |
| 1 | 537.67 | 800.7 | 267 | 245.8 | 20.8 | 8.55 |
| 1 | 551.31 | 415.3 | 406 | 93.7 | 29.3 | 0 |
| 1 | 66.49 | 529.1 | 254 | 392 | 76.1 | 12.41 |
| 1 | 2347.63 | 653.6 | 655 | 201.1 | 192.6 | 50.45 |
| 1 | 62.33 | 290.8 | 187 | 81.6 | 75.1 | 35.37 |
| 1 | 36.13 | 584.9 | 303 | 189.7 | 69.8 | 51.4 |
| 1 | 698.12 | 304 | 460 | 81.8 | 16.8 | 17.21 |
| 1 | 67.8 | 388.3 | 281 | 137 | 40 | 4.34 |
| 1 | 105.13 | 405.9 | 746 | 360 | 114.2 | 32.07 |
| 1 | 2163.19 | 684.3 | 447 | 152 | 61.4 | 32.69 |
| 1 | 46.64 | 446 | 395 | 237.2 | 105.1 | 33.52 |
| 1 | 120.29 | 235 | 642 | 140.1 | 46.8 | 12.23 |
| 1 | 48.65 | 293.8 | 223 | 94.6 | 27.8 | 6.82 |
| 1 | 36.36 | 745.5 | 467 | 109.5 | 29 | 7.46 |
| 1 | 729.93 | 1973.9 | 2370 | 734.3 | 116.8 | 12.96 |
| 1 | 224.96 | 557.3 | 1504 | 1282 | 257.9 | 97.02 |
| 1 | 304.35 | 716.4 | 567 | 1148.2 | 288.4 | 88.07 |
| 1 | 75.67 | 556.2 | 470 | 542.7 | 708.6 | 172.21 |
| 1 | 4.18 | 96.8 | 683 | 556.5 | 152.8 | 173.23 |

Table 6.3.1.2 Plaice in 7.d. Landings Residuals

| AGE | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | -0.7317623 | 0.7566031 | -0.4537843 | -0.2965485 | 0.2433267 | -0.0335535 | -0.1074725 |
| 1981 | -1.3342082 | 0.2234025 | 0.3884756 | 0.3356951 | -0.1290887 | -0.1795775 | 0.2904463 |
| 1982 | 0.656489 | 0.2266889 | -0.0949777 | -0.0819194 | 0.0580407 | 0.3417678 | -0.3046804 |
| 1983 | -0.4874645 | 0.193992 | -0.3108887 | 0.092499 | -0.3021358 | -0.203802 | -0.0610599 |
| 1984 | 0.6120163 | -0.4227322 | -0.1980962 | 0.1460666 | 0.1639283 | 0.1455761 | 0.2762606 |
| 1985 | -0.5629748 | 0.5532505 | -0.2708043 | 0.2638308 | -0.5253418 | -0.0645073 | 0.004894 |
| 1986 | 0.564304 | 0.3181647 | -0.0784105 | 0.1795455 | 0.1592601 | 0.5011414 | -0.5382334 |
| 1987 | -1.8573791 | 0.4775731 | -0.1468664 | 0.0219268 | 0.0649163 | -0.3826158 | 0.2550488 |
| 1988 | -1.9362587 | 0.5622501 | 0.1174346 | -0.0546161 | -0.1904098 | -0.006112 | 0.1022216 |
| 1989 | 1.4967842 | 0.3512363 | -0.381461 | -0.1915061 | 0.3214279 | 0.0243125 | -0.0475156 |
| 1990 | 0.9707152 | -0.0580355 | 0.2106991 | -0.0707438 | -0.1352495 | -0.037476 | 0.1482987 |
| 1991 | -0.4267999 | 0.4580829 | 0.2665366 | 0.2942429 | 0.0548876 | -0.0771983 | -0.1433432 |
| 1992 | -0.6884845 | 0.0970929 | 0.32322 | 0.0181352 | -0.0345705 | 0.0667483 | 0.0146508 |
| 1993 | -0.554017 | 0.2589953 | -0.1474135 | -0.0869141 | -0.183979 | -0.09832 | -0.1391778 |
| 1994 | 0.3271496 | 0.2645401 | 0.2026866 | 0.3248958 | 0.3548035 | 0.2960378 | 0.168455 |
| 1995 | 0.6418369 | 0.5444338 | -0.113989 | -0.1056738 | -0.1372602 | -0.3433138 | -0.1683983 |
| 1996 | -0.1084539 | 0.1492848 | 0.1480712 | -0.177267 | 0.0116047 | -0.061425 | -0.1399089 |
| 1997 | -1.0029445 | 0.1346697 | 0.3188984 | 0.7181046 | 0.491243 | 0.3930009 | 0.2718472 |
| 1998 | -0.1416657 | -0.1406508 | 0.438704 | 0.0357727 | -0.0100945 | -0.2434586 | $-0.2862662$ |
| 1999 | 0.3337157 | -0.3589311 | 0.1752386 | 0.4552492 | 0.0199685 | 0.3350789 | 0.0663487 |
| 2000 | 0.1027707 | 0.6558747 | -0.3740824 | -0.270096 | -0.1172459 | -0.1161922 | 0.0812366 |
| 2001 | 0.4033466 | -0.2864938 | -0.2256259 | -0.4253859 | -0.1481753 | -0.227334 | -0.0580562 |
| 2002 | -0.45984 | 0.4543747 | 0.0182458 | 0.0827547 | 0.6207151 | -0.0086562 | 0.0951584 |
| 2003 | -5.1954149 | -0.0223658 | -0.5396951 | 0.2000351 | -0.1894234 | 0.0299125 | 0.1707002 |
| 2004 | 2.6741098 | 0.8517705 | -0.2293421 | -0.3987292 | -0.0474681 | -0.1843397 | -0.2454659 |
| 2005 | 0.5785501 | 0.662106 | -0.4892391 | -0.1555769 | 0.0535073 | 0.0429811 | -0.0076851 |
| 2006 | 0.6946995 | 0.5427154 | -0.4395654 | 0.1117793 | 0.197065 | -0.2228395 | 0.0757119 |
| 2007 | 0.7945855 | 0.3292781 | -0.5980723 | -0.3150248 | 0.1285787 | 0.053215 | 0.0347295 |
| 2008 | -0.1588189 | 0.3007986 | -0.4355213 | -0.1080558 | 0.1208411 | -0.0346606 | -0.2510059 |
| 2009 | 0.2858867 | -0.1065306 | -0.3539347 | -0.1597542 | 0.0746052 | -0.0553035 | -0.2098879 |
| 2010 | 0.3328369 | 0.0897623 | -0.4613077 | 0.2484095 | 0.2786715 | 0.4786274 | 0.0363171 |
| 2011 | -1.4481915 | 0.0072165 | -0.6669212 | -0.514388 | -0.1009961 | -0.6833003 | -0.5897582 |
| 2012 | -0.0079514 | -0.022589 | -0.0711335 | 0.0119617 | 0.0574442 | 0.4348787 | -0.0366851 |
| 2013 | -0.0597915 | 0.0724931 | -0.1075138 | 0.0076288 | 0.1831061 | 0.2289692 | -0.2414452 |
| 2014 | -0.0200016 | 0.0153136 | 0.4397213 | 0.0495386 | 0.1542901 | 0.1470297 | -0.4317627 |
| 2015 | -0.0762711 | -0.1305248 | -0.0294269 | -0.1150904 | -0.1566889 | 0.0556973 | -0.0255026 |

Table 6.3.1.2 (cont.) Plaice in 7.d. Discards Residuals

|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | -0.00908802 | 0.00031716 | -0.08771148 | -0.82880826 | -0.7855614 | 0.08560163 | 0.42011066 |
| 2007 | -0.12960469 | 0.50044144 | -0.01494139 | 0.39017146 | -2.63289603 | 0.09618416 | 0.57302124 |
| 2008 | 0.16434532 | -0.42821622 | -0.94066635 | 0.16185145 | 0.92037458 | 1.65224772 | 0.94337154 |
| 2009 | -0.06134169 | 0.09480914 | 0.07303661 | 0.03767075 | 0.50453007 | 2.60064971 | 1.8307582 |
| 2010 | -0.09358731 | 0.66840533 | 0.37618488 | 0.9739187 | 1.17859469 | 1.12768993 | 1.76332543 |
| 2011 | 0.05630504 | 0.2552561 | -0.61659755 | -0.13914334 | 0.37563609 | 3.03444045 | 1.41390177 |
| 2012 | 0.01264133 | -0.02219619 | -0.07090686 | 0.01297867 | 0.07705162 | 1.56638513 | 3.97886867 |
| 2013 | -0.02371232 | 0.07318064 | -0.10730794 | 0.00829163 | 0.19462398 | 0.33090886 | 0.52945203 |
| 2014 | -0.01485282 | 0.01597445 | 0.44002571 | 0.04992961 | 0.15651739 | 0.16043345 | -0.33074677 |
| 2015 | 0.02012124 | -0.12977373 | -0.02897518 | -0.11415927 | -0.15517922 | 0.05971473 | -0.0003182 |

Table 6.3.1.3 Plaice in 7.d. Survey residuals

| UK BTS |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |
| age | 1 | 2 |  |  | 6 |  |
| 1989 | -1.43148745 | -0.62624065 | -0.02135472 | 0.38638918 | -0.07116134 | 0.13027007 |
| 1990 | -0.53547634 | -0.61966855 | -0.45519266 | -0.15493513 | 0.17307268 | -0.32496766 |
| 1991 | -0.37975962 | -0.01601455 | 0.17993733 | 0.05080842 | 0.19403026 | -0.19334254 |
| 1992 | -0.25173593 | 0.01525521 | -0.13088954 | 0.38830939 | 0.4711237 | 0.30405314 |
| 1993 | -1.10447529 | -0.21383867 | -0.28839183 | -0.21891379 | 0.16725896 | -0.06463491 |
| 1994 | -0.29294124 | -0.40667679 | -0.23559149 | 0.2023231 | 0.07457078 | -0.69917606 |
| 1995 | -0.43951373 | -0.62824213 | -0.49359842 | 0.06902363 | -0.04921443 | -0.20408029 |
| 1996 | -0.14796569 | -0.2428682 | -1.14305435 | -0.74808974 | -0.11848734 | 0.12857335 |
| 1997 | 0.06546143 | 0.04718776 | -0.05637104 | -0.44775196 | 0.53670304 | -0.5801213 |
| 1998 | 0.34397762 | -0.30119526 | -0.29034675 | 0.08261055 | -0.16197706 | 0.42278586 |
| 1999 | -0.36692911 | -0.83552111 | -0.10786985 | -0.41604123 | -0.19430781 | -0.02637398 |
| 2000 | -0.14318957 | 0.44197261 | 0.25354944 | 0.32081673 | 0.36853388 | 0.18878627 |
| 2001 | 0.31586228 | 0.0164717 | 0.36631755 | 0.39036214 | 0.64879427 | 0.34510529 |
| 2002 | 0.2890561 | 0.36695719 | 0.32498023 | 0.25111136 | -0.27776436 | 0.16742688 |
| 2003 | -0.35590819 | 0.26018649 | 0.10035136 | 0.13234092 | 0.0286 | -0.49352247 |
| 2004 | 1.01125291 | 0.18404982 | 0.13944107 | 0.16901792 | -0.51335679 | -1.21151257 |
| 2005 | -0.08409085 | 0.40893837 | 0.08348681 | 0.14898737 | 0.40269139 | 0.01515998 |
| 2006 | 0.60743401 | 0.3347427 | 0.11918789 | 0.03029914 | -0.09382152 | 0.13105584 |
| 2007 | 0.94953223 | 0.43026296 | 0.00034771 | -0.16285356 | -0.00967448 | 0.06150179 |
| 2008 | 0.34571173 | 0.5625445 | -0.01437182 | -0.32238223 | -0.0865295 | -0.04344882 |
| 2009 | 0.56902137 | 0.19737558 | 0.44053244 | 0.10107347 | -0.07526721 | -0.11131366 |
| 2010 | -0.08160808 | 0.19337409 | -0.1395587 | 0.27676884 | -0.01037471 | -0.21412174 |
| 2011 | 0.20545346 | 0.24716936 | 0.07082351 | -0.07531371 | 0.49768285 | 0.34072371 |
| 2012 | -0.64787945 | -0.39057798 | -0.14675079 | -0.12537556 | -0.47304607 | 0.18974073 |
| 2013 | -0.42180968 | 0.00502969 | -0.04178826 | 0.15013855 | -0.25475869 | -0.2598642 |
| 2014 | 0.92208764 | 0.91618134 | 0.07301929 | 0.10070953 | 0.09262729 | 0.04141455 |
| 2015 | 0.3689431 | 0.91530687 | 0.26546319 | 0.03628505 | -0.05467066 | 0.00813469 |
|  |  |  |  |  |  |  |

Table 6.3.1.3 (cont.) Plaice in 7.d. Survey Residuals

| FR GFS |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1 | 2 | 3 | 4 | 5 | 6 |
| 1993 | 1.353593 | 0.0210181 | 0.0548124 | -0.0089883 | -0.612795 | -0.5389702 |
| 1994 | 0.5411788 | -0.1679686 | -0.717743 | -0.4907609 | -0.1001986 | 0.7926329 |
| 1995 | -0.077728 | 0.8099477 | 0.7236571 | 0.7416236 | 0.4452086 | 2.0290029 |
| 1996 | -0.5347635 | -0.5160409 | -0.7795479 | -0.3547855 | -0.0282261 | 0.9151427 |
| 1997 | 0.073288 | -0.3668155 | 0.6459888 | 0.1960825 | 0.3875457 | 1.7552893 |
| 1998 | 0.1434398 | -0.6015105 | -0.6620311 | -0.2870276 | -1.1096102 | 1.2627494 |
| 1999 | 0.5408585 | -0.3405981 | 0.0625838 | 0.1792969 | -0.0923184 | 1.6330874 |
| 2000 | 0.5781173 | 0.8820428 | 0.2112203 | 0.3447399 | 0.6531712 | 1.0170063 |
| 2001 | -0.098032 | -0.3372665 | -0.2423324 | 0.2072461 | -0.492258 | -0.2842516 |
| 2002 | 0.2373909 | 0.2318073 | 0.6757749 | 0.5673839 | 0.6286002 | -0.1522591 |
| 2003 | 0.1732396 | 0.0508293 | -0.1632947 | -0.6841587 | 0.0525786 | 0.2300011 |
| 2004 | 0.179822 | 0.0174283 | -0.3888267 | 0.0795215 | -0.9397006 | 1.1059182 |
| 2005 | 0.3427379 | 0.7374616 | 0.9816194 | 0.3378208 | 0.6768258 | 1.0192598 |
| 2006 | 0.9064749 | 0.39293 | -0.0978014 | 0.1265211 | -0.0892722 | 0.6352713 |
| 2007 | 0.2594098 | 0.3835484 | 0.5537636 | 0.4075869 | 0.2970984 | -0.6784026 |
| 2008 | -0.6873221 | 0.6642244 | 0.166533 | -0.1824748 | -0.8778031 | 0.2159958 |
| 2009 | -0.9186273 | -0.7361389 | -0.4567317 | -0.5629428 | -1.175137 | -0.2468969 |
| 2010 | -0.4549106 | -0.5149893 | -0.7538173 | -0.7906015 | -1.0126621 | 0.0176862 |
| 2011 | 0.2211274 | 0.597396 | 0.5175538 | 0.0763692 | -0.8416741 | 1.5953954 |
| 2012 | -0.3587013 | -0.1774395 | 0.4899349 | 0.2131379 | 0.5573373 | 0.3428993 |
| 2013 | -0.1473372 | -0.4719367 | 0.0211648 | -0.2817425 | -0.1847234 | -0.5596347 |
| 2014 | -0.5533869 | -0.6841747 | -0.0623705 | 0.2397628 | -0.1292369 | -0.7164262 |
| 2015 | -1.487909 | -0.4383178 | -0.0667455 | -0.6293846 | -0.4904655 | -0.6884282 |

Table 6.3.1.4 Plaice in 7.d. Fishing mortality (F) at age

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.0138004 | 0.13028 | 0.400872 | 0.311596 | 0.170516 | 0.0990706 | 0.0990706 |
| 1981 | 0.0151641 | 0.135083 | 0.437482 | 0.397358 | 0.238643 | 0.139975 | 0.139975 |
| 1982 | 0.0171941 | 0.147522 | 0.485567 | 0.474815 | 0.30375 | 0.185111 | 0.185111 |
| 1983 | 0.020738 | 0.178502 | 0.557166 | 0.498611 | 0.320219 | 0.21464 | 0.21464 |
| 1984 | 0.0254172 | 0.230258 | 0.646873 | 0.460697 | 0.281243 | 0.216699 | 0.216699 |
| 1985 | 0.0266535 | 0.262725 | 0.69836 | 0.417941 | 0.244133 | 0.208412 | 0.208412 |
| 1986 | 0.0204471 | 0.221842 | 0.64647 | 0.416047 | 0.248993 | 0.209747 | 0.209747 |
| 1987 | 0.0144166 | 0.158761 | 0.539217 | 0.449105 | 0.293379 | 0.225642 | 0.225642 |
| 1988 | 0.0155992 | 0.139092 | 0.469054 | 0.474335 | 0.341244 | 0.251207 | 0.251207 |
| 1989 | 0.0371468 | 0.195869 | 0.476332 | 0.451333 | 0.343745 | 0.278126 | 0.278126 |
| 1990 | 0.125351 | 0.349958 | 0.528068 | 0.398773 | 0.30541 | 0.285281 | 0.285281 |
| 1991 | 0.300703 | 0.534324 | 0.565018 | 0.349383 | 0.255481 | 0.250255 | 0.250255 |
| 1992 | 0.357016 | 0.567657 | 0.552624 | 0.322504 | 0.21723 | 0.190575 | 0.190575 |
| 1993 | 0.285807 | 0.503979 | 0.538842 | 0.329737 | 0.20742 | 0.155333 | 0.155333 |
| 1994 | 0.224973 | 0.466191 | 0.577294 | 0.390106 | 0.244557 | 0.168853 | 0.168853 |
| 1995 | 0.193077 | 0.459784 | 0.657503 | 0.505823 | 0.338275 | 0.236151 | 0.236151 |
| 1996 | 0.156364 | 0.413831 | 0.684624 | 0.623074 | 0.456956 | 0.325213 | 0.325213 |
| 1997 | 0.108013 | 0.303888 | 0.580475 | 0.645404 | 0.511905 | 0.348459 | 0.348459 |
| 1998 | 0.0803795 | 0.223953 | 0.458886 | 0.58061 | 0.472692 | 0.296764 | 0.296764 |
| 1999 | 0.0927587 | 0.230864 | 0.427367 | 0.497773 | 0.380464 | 0.22703 | 0.22703 |
| 2000 | 0.17698 | 0.370834 | 0.522322 | 0.436218 | 0.286588 | 0.1753 | 0.1753 |
| 2001 | 0.277201 | 0.584372 | 0.677027 | 0.395369 | 0.225355 | 0.151152 | 0.151152 |
| 2002 | 0.162058 | 0.530115 | 0.720835 | 0.371445 | 0.206725 | 0.159852 | 0.159852 |
| 2003 | 0.0386855 | 0.274851 | 0.605675 | 0.358628 | 0.220226 | 0.19884 | 0.19884 |
| 2004 | 0.0108886 | 0.144628 | 0.489654 | 0.348596 | 0.238685 | 0.240121 | 0.240121 |
| 2005 | 0.0094362 | 0.130815 | 0.458758 | 0.334538 | 0.232605 | 0.236526 | 0.236526 |
| 2006 | 0.019186 | 0.183017 | 0.491297 | 0.315021 | 0.204328 | 0.196747 | 0.196747 |
| 2007 | 0.0382231 | 0.261727 | 0.538045 | 0.291185 | 0.172489 | 0.15843 | 0.15843 |
| 2008 | 0.0402568 | 0.281089 | 0.54885 | 0.264552 | 0.147148 | 0.135287 | 0.135287 |
| 2009 | 0.0276595 | 0.23487 | 0.503962 | 0.237242 | 0.127529 | 0.115886 | 0.115886 |
| 2010 | 0.0179739 | 0.169381 | 0.40778 | 0.211011 | 0.11191 | 0.0917004 | 0.0917004 |
| 2011 | 0.0141284 | 0.118259 | 0.299842 | 0.187506 | 0.100218 | 0.0679323 | 0.0679323 |
| 2012 | 0.0145018 | 0.0910184 | 0.221664 | 0.168244 | 0.0937239 | 0.0543524 | 0.0543524 |
| 2013 | 0.0205087 | 0.0869097 | 0.181579 | 0.153926 | 0.0934989 | 0.0539198 | 0.0539198 |
| 2014 | 0.037611 | 0.101204 | 0.165418 | 0.143336 | 0.0985855 | 0.0655779 | 0.0655779 |
| 2015 | 0.0785829 | 0.130216 | 0.15898 | 0.134663 | 0.106881 | 0.088364 | 0.088364 |

Table 6.3.1.4 Plaice in 7.d. Stock number from the assessment.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 66487.2 | 30226.8 | 9931.93 | 2438.46 | 1978.16 | 670.93 | 1875.38 |
| 1981 | 34816.7 | 46067.5 | 18640.7 | 4672.91 | 1254.42 | 1171.81 | 1620.07 |
| 1982 | 65693.4 | 24090.9 | 28273.5 | 8455.08 | 2206.32 | 694.15 | 1705.13 |
| 1983 | 59878.4 | 45363.3 | 14602.8 | 12222.2 | 3694.53 | 1143.94 | 1400.68 |
| 1984 | 61544.2 | 41201.6 | 26658.4 | 5876.41 | 5215.04 | 1884.26 | 1442.3 |
| 1985 | 77470 | 42150.2 | 22991.4 | 9807.33 | 2604.26 | 2765.45 | 1881.63 |
| 1986 | 155094 | 52991.8 | 22769.3 | 8033.81 | 4536.2 | 1433.21 | 2650.44 |
| 1987 | 94476 | 106749 | 29820.5 | 8379.96 | 3722.93 | 2484.31 | 2325.99 |
| 1988 | 61809.7 | 65420.1 | 63983.4 | 12217.6 | 3757.07 | 1950.4 | 2696.67 |
| 1989 | 41183.8 | 42749.6 | 39990.3 | 28119.7 | 5341.16 | 1876.29 | 2539.41 |
| 1990 | 42485.5 | 27876.9 | 24689.8 | 17447.6 | 12579.1 | 2660.72 | 2348.88 |
| 1991 | 73290.6 | 26330 | 13801 | 10229 | 8226.26 | 6511.22 | 2645.81 |
| 1992 | 88777.8 | 38115.8 | 10840.5 | 5510.32 | 5066.96 | 4476.09 | 5008.65 |
| 1993 | 41775.9 | 43642.1 | 15178.4 | 4382.25 | 2803.92 | 2864.55 | 5506.95 |
| 1994 | 35888.6 | 22052.2 | 18521.7 | 6220.99 | 2213.83 | 1600.79 | 5034.93 |
| 1995 | 62768.2 | 20132.7 | 9719.39 | 7304.92 | 2958.62 | 1217.83 | 3937.38 |
| 1996 | 69714.6 | 36352.8 | 8930.41 | 3537.85 | 3094.5 | 1481.94 | 2859.81 |
| 1997 | 118778 | 41885.8 | 16883.6 | 3163.69 | 1332.88 | 1376.53 | 2203.32 |
| 1998 | 56203.8 | 74899.5 | 21714 | 6637.7 | 1165.6 | 561.21 | 1774.93 |
| 1999 | 48589.6 | 36434.1 | 42059.9 | 9640.48 | 2609.23 | 510.4 | 1219.74 |
| 2000 | 62202.8 | 31110.7 | 20318.7 | 19271.5 | 4116.89 | 1252.94 | 968.58 |
| 2001 | 59642.8 | 36610 | 15083.7 | 8466.51 | 8752.24 | 2171.48 | 1309.69 |
| 2002 | 77267.3 | 31755.7 | 14337.1 | 5384.31 | 4005.42 | 4907.94 | 2102.48 |
| 2003 | 38573.7 | 46159.9 | 13129.5 | 4898.43 | 2608.94 | 2288.33 | 4197.32 |
| 2004 | 47696.9 | 26070 | 24634.8 | 5033.34 | 2404.13 | 1470.52 | 3734.64 |
| 2005 | 42294.8 | 33144.5 | 15848.2 | 10605.9 | 2495.24 | 1330.29 | 2876.09 |
| 2006 | 40842.5 | 29433.3 | 20429.1 | 7037.11 | 5332.23 | 1389.14 | 2332.58 |
| 2007 | 51775.4 | 28146.9 | 17219 | 8780.79 | 3607.73 | 3053.66 | 2147.58 |
| 2008 | 70735 | 35008.5 | 15220 | 7062.97 | 4610.25 | 2132.91 | 3118.54 |
| 2009 | 110054 | 47731.1 | 18567.3 | 6175.94 | 3808.42 | 2795.56 | 3222.37 |
| 2010 | 173358 | 75204.9 | 26512.4 | 7880.11 | 3422.32 | 2355.1 | 3765.04 |
| 2011 | 232267 | 119615 | 44600 | 12388.1 | 4482.72 | 2149.66 | 3922.72 |
| 2012 | 117472 | 160880 | 74658.4 | 23214.9 | 7214.74 | 2848.84 | 3985.72 |
| 2013 | 122771 | 81336.9 | 103187 | 42020.6 | 13783.2 | 4614.96 | 4547.33 |
| 2014 | 145350 | 84496.6 | 52383.4 | 60452.6 | 25308.3 | 8818.48 | 6098.7 |
| 2015 | 67953 | 98340.4 | 53646 | 31189.2 | 36797.3 | 16110.1 | 9814.25 |

Table 6.3.1.6 Plaice in 7.d. Summary table (Outputs from the model)

|  | Recruitment | SSB | Catch | LANDINGS | Total Biomass | Fbar |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 66487.2 | 8180.7 | 2292 | 1826 | 28372 | 0.24551 |
| 1981 | 34816.7 | 10781 | 3814 | 3074 | 29819 | 0.30336 |
| 1982 | 65693.4 | 13150 | 5260 | 4444 | 32644 | 0.36231 |
| 1983 | 59878.4 | 13256 | 5339 | 4386 | 33028 | 0.39766 |
| 1984 | 61544.2 | 13273 | 6160 | 4893 | 33303 | 0.40138 |
| 1985 | 77470 | 13237 | 6144 | 4783 | 34214 | 0.39221 |
| 1986 | 155094 | 13121 | 5965 | 4514 | 44787 | 0.38031 |
| 1987 | 94476 | 15623 | 6920 | 5168 | 49410 | 0.37684 |
| 1988 | 61809.7 | 20628 | 8921 | 7190 | 47644 | 0.38396 |
| 1989 | 41183.8 | 22361 | 8804 | 7347 | 39909 | 0.38738 |
| 1990 | 42485.5 | 19277 | 7418 | 5952 | 32203 | 0.37938 |
| 1991 | 73290.6 | 15254 | 6989 | 4597 | 29840 | 0.35503 |
| 1992 | 88777.8 | 12404 | 7138 | 3916 | 30215 | 0.32073 |
| 1993 | 41775.9 | 11262 | 5778 | 3567 | 25238 | 0.30783 |
| 1994 | 35888.6 | 10631 | 4838 | 3391 | 20947 | 0.3452 |
| 1995 | 62768.2 | 9206.1 | 4797 | 3267 | 21283 | 0.43444 |
| 1996 | 69714.6 | 7848.7 | 5156 | 3382 | 23498 | 0.52247 |
| 1997 | 118778 | 7922.4 | 5486 | 3552 | 31773 | 0.52156 |
| 1998 | 56203.8 | 10351 | 5768 | 4043 | 33533 | 0.45224 |
| 1999 | 48589.6 | 14138 | 6629 | 5137 | 32941 | 0.38316 |
| 2000 | 62202.8 | 15533 | 7182 | 5343 | 32473 | 0.35511 |
| 2001 | 59642.8 | 14013 | 7421 | 4921 | 31052 | 0.36223 |
| 2002 | 77267.3 | 12501 | 6248 | 4198 | 31158 | 0.36471 |
| 2003 | 38573.7 | 11918 | 4735 | 3587 | 28297 | 0.34584 |
| 2004 | 47696.9 | 12561 | 4449 | 3747 | 28343 | 0.32926 |
| 2005 | 42294.8 | 13136 | 4068 | 3459 | 28496 | 0.31561 |
| 2006 | 40842.5 | 13652 | 4480 | 3705 | 28902 | 0.30185 |
| 2007 | 51775.4 | 13865 | 4606 | 3676 | 30014 | 0.29004 |
| 2008 | 70735 | 13691 | 4734 | 3476 | 33613 | 0.27396 |
| 2009 | 110054 | 14375 | 4912 | 3561 | 42916 | 0.24615 |
| 2010 | 173358 | 17278 | 5374 | 3933 | 61236 | 0.2056 |
| 2011 | 232267 | 24231 | 6530 | 4702 | 87194 | 0.16387 |
| 2012 | 117472 | 36661 | 6118 | 2977 | 95269 | 0.1345 |
| 2013 | 122771 | 49924 | 6400 | 3674 | 96915 | 0.12073 |
| 2014 | 145350 | 55810 | 6134 | 3288 | 97071 | 0.11823 |
| 2015 | 67953 | 54378 | 5788 | 3253 | 86361 | 0.12222 |

Table 6.5.3.1.1 Plaice in 7.d. Management options for 2016 and their effects on the resident stock.

| Variable | Value | Source | Notes |
| :--- | :--- | :--- | :--- |
| F ages 3-6 (2016) | 0.27 | AAP | Assuming that the 7d proportion of the TAC <br> 2016 is fully landed |
| SSB (2017) | 61116 | AAP | Short term forecast (STF), tonnes |
| Rage1 (2017) | 155235 | GM 2010-2013 | Thousands individuals |
| Rage1 (2017) | 155235 | GM 2010-2013 | Thousands individuals |
| Catch (2016) | 14074 | AAP | t (resident stock) |
| Landings (2016) | 8223 | AAP | t (resident stock) |
| Discards (2016) | 5851 | AAP | projection based on the 2013-2015 discard <br> ratio (by age) |


| Rationale | BASIS | $\begin{aligned} & \text { CATCH } \\ & (2017) \end{aligned}$ | Wanted CATCH (2017) | UnWANTED CATCH (2017) | $\begin{gathered} \text { F } \\ \text { TOTAL } \\ \text { CATCH } \\ (2017) \end{gathered}$ | $\begin{aligned} & \text { \%SSB CHANGE } \\ & (2018 / 2017) \end{aligned}$ | $\begin{aligned} & \text { \% CHANGE IN } \\ & \text { WANTED } \\ & \text { CATCH } \\ & (2017 / 2015) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MSY <br> Approach | Fmsy | 12805 | 7550 | 5255 | 0.25 | -3\% | +155\% |
| Other options | Fpa | 17607 | 10402 | 7205 | 0.36 | -12\% | +252 |
|  | Fsq (Fbar 2015) | 6502 | 3824 | 2678 | 0.12 | +9\% | +29\% |
|  | Landings 2015 roll over | 5030 | 2957 | 2073 | 0.09 | +12\% | +0\% |
|  | Landings $2015+$ <br> 20\% | 6034 | 3548 | 2486 | 0.11 | +10\% | +20\% |
|  | Landings $2015-20 \%$ | 4025 | 2365 | 1660 | 0.07 | +14\% | -20\% |



Figure 6.2.1.1. Plaice in 7.d. Official landings.


Figure 6.2.2.1: Plaice in 7.d. Proportions of total landings per country with and without age distribution provided.


Figure 6.2.3.1. Plaice in 7.d. Age composition of the landings.


Figure 6.2.3.2. Plaice in 7.d. Age composition of the discards.

## Landings



## Discards



Figure 6.2.1.2. Plaice in 7.d. 2015 Age distribution in the sampled landings and discards per quarter. (Number of individuals)


Figure 6.2.1.3. Plaice in 7.d. Discards at age ratio (discards numbers/landings numbers) per age and through time.



Figure 6.2.3.1. Plaice in 7.d. Stock, Catch weight and discard weight.


Figure 6.2.6.1 Plaice in 7.d. Consistency between surveys: Mean standardised indices by surveys for each age.



Figure 6.2.6.2: Old and new UK BTS and FR GFS indices.


Figure 6.2.6.3: Old and new UK BTS and FR GFS indices consistencies


Figure 6.2.6.4: Effect of the changes in the survey indices on the 2015 assessment. Red: assessment with old indices, black: assessment with new indices.


Figure 6.3.1.1. Plaice in 7.d. Landings (left) and discards (right) time series: observed (dots) vs modelled (line), and per age (from 1 to 6: bottom panels).


Figure 6.3.1.2. Plaice in 7.d. Survey residuals from the AAP assessment.


Figure 6.3.1.3. Plaice in 7.d. Summary of assessment results.


Figure 6.3.1.4. Plaice in 7.d. Retrospective patterns.


Figure 6.3.1.5. Plaice in 7.d. Estimated stock numbers.


Figure 6.5.2.1 Plaice in 7.d. Time series of (left) proportion of the catch taken in 7.e over the total catch for $7 . \mathrm{d}-\mathrm{e}$, and (right) proportion of the catch of fish coming from . 7 e and 4 over the $7 . \mathrm{d}$ catch, and the averages used.

## 7 Plaice (Pleuronectes platessa) in Subdivision 3.aN (Skagerrak)

The plaice in Skagerrak has been benchmarked in February 2015 (ICES WKPLE), and is now assessed together with the North Sea Plaice (see details in ICES WGNSSK 2015 report). All information related to the assessment of the combined stock is found in section 8. But the current section on Plaice in Skagerrak is maintained to display the relevant indicators.

The existence of a resident plaice population in Skagerrak, with a distinct genetic print has been demonstrated (Ulrich et al., in press). However, this population is importantly mixed with the North Sea population, also during spawning season, indicating that the Skagerrak area also belongs to the natural distribution area of North Sea plaice.

During summer, there is likely an important inflow from the North Sea population, entering Skagerrak to feed. This inflow has increased over the recent years, consistently with the increase of abundance of the North Sea stock. The largest part of the fishery occurs in this period, and in the most westerly part of the Skagerrak close to the North Sea border (Jammer Bay). Therefore, much (and likely most) of the commercial catches recorded for Skagerrak may belong to the North Sea component, although the geographical patterns of mixing with the local Skagerrak component is not known in detail at present.

There is thus no scope for a stock assessment of the Skagerrak population alone at present. Nevertheless, WGNSSK suggested a number of indicators to monitor fisheries trends in the Skagerrak. This routine scrutiny could potentially detect a departure from the current situation and an increased risk of local depletion of the resident population. In the medium-term, other actions could be undertaken to improve the monitoring of the local component in Skagerrak. Such actions have not been launched at present, but could motivate future research efforts.

### 7.1 Ecosystem aspects

### 7.1.1 Fisheries

## Technical Conservation Measures

Minimum Landing Size is 27 cm .

## Changes in fleet dynamics

A detailed description of the fishing activities in area 3.a is available in STECF (2015) ${ }^{1}$, separated by area (Skagerrak= area 3.b1).

In 2015, the share of landings coming from Fully Documented Fisheries in the Skagerrak reduced sensibly compared with previous years, down to 586 t (against 1500 tonnes

1 https://stecf.jrc.ec.europa.eu/documents/43805/1040968/2015-07_STECF+15-12++FDI+2015_JRC97365.pdf
and http://datacollection.jrc.ec.europa.eu/dd/effort/graphs
in 2014), corresponding to around $6 \%$ of total landings in the area ( $17 \%$ in 2014). (InterCatch data). There are less Danish landings operating under that scheme, but some Dutch flyshooters have been recorded in FDF.

### 7.1.2 Management

According to the agreed records of EU-Norway negotiations in December 2015, the North Sea plaice long-term management plan was not used as the basis for TAC for the combined stock. An interim agreement was reached to keep the TAC in 2016 at the 2015 level in both areas. These TACs were adjusted by $2.6 \%$ in the North Sea and by $17 \%$ in the Skagerrak in order to take into account the inclusion of plaice (trawls and beam trawls with mesh size equal to or greater than 100 mm ) in the EU landing obligation, resulting in a TAC of 11766 tonnes in the Skagerrak.

### 7.2 Data available

### 7.2.1 Catch

The official landings reported to ICES are not distinguished between Skagerrak (3.asN) and Kattegat (3.aS) in the FAO areas definitions used by Eurostat, so this information is not presented.

The annual landings used by the Working Group, available since 1972, are given by country in Tables 7.1.1. Denmark stands for the largest part of landings ( $80 \%$ in 2015). Misreporting is not considered an issue.

As in previous years, information was provided by DCF metier as specified in the data call and InterCatch was used to raise catch-at-age information.

Age information is provided by Denmark, and discards information is provided by Denmark, Sweden and Germany (Tables 7.3.1 and 7.3.2)

The small issue in the older Swedish InterCatchdata discovered during WGNSSK 2015 will still have to be reprocessed during the benchmark next year.

For 2015, landings strata for which discards ratios are available summed up to $78 \%$ of all landings weight (Figure 7.3.1). Discards raising was performed as in previous years, by grouping all static gears together, small mesh size fishery $32-69 \mathrm{~mm}$ together, and demersal fisheries $>90 \mathrm{~mm}$ together. Age information was used to raise to international catches all metiers together, on a quarter basis for landings and on a yearly basis for discards.

Overall 2015 discards were estimated at 677 tonnes, corresponding to a discards ratio to catches of $6,4 \%$, which is lower than estimations from previous years.

### 7.2.2 Weight at age

Weight at age in landings is presented in Table 7.3.4 and Figure 7.3.2.

### 7.2.3 Catch, effort and research vessel data

Landings and effort data by gear were computed at the level of the ICES rectangle. 2002-2014 Data for the European fleet was available from the STECF online data [http://datacollection.jrc.ec.europa.eu/dd/effort/graphs], on the basis of the main "cod plan categories". Provisional 2015 data from Sweden and Denmark were added in the same format.

IBTS data are available in the area 3.aN. Since 2007 the WG discussed the limited spatial coverage of the surveys with regards to main fishing grounds. The IBTS sampling in Skagerrak has only few hauls in the Western Skagerrak. Since 2014, the Danish part of IBTS has added one haul in Skagerrak, located more coastally (43F9) than usual IBTS hauls, and thus more directly into the expected plaice distribution area. In 2014, this haul yielded the most important density ever recorded in the Skagerrak, and WKPLE suggested to remove that haul from the index until more coastal hauls were taken in order to disentangle possible effects of depth from just an outlier. In 2015 and 2016, hauls in shallow waters were also performed and obtained also a high CPUE, though less extreme than in 2014 (Figure 7.3.5).

### 7.3 Data analyses

### 7.3.1 Catch-at-age matrix

There are almost no landings from age 1 plaice, and generally poor tracking of the cohorts in the landings. (Figure 7.3.2). There has been a shift in the age distribution in 2003, from predominantly age $4-5$ to now mainly age $3-4$. Weight at age has been increasing over the decade for the main ages. Discards are mainly on ages 2 and 3 (Figure 7.3.3)

### 7.3.2 Spatial information on catches

Information of commercial catches is still presented accordingly for the purpose of the monitoring of the situation (Figure 7.4.1). Nearly all catches are taken in the Western area ( $\sim 98 \%$ in 2015), while plaice by-catches in the targeted Nephrops fishery in the Eastern area have dropped to very low levels with the increased adoption of more selective gears.

Catches trends in fishing patterns by the main gear groups across both areas were computed. 2015 lpue for the main fishery (TR1 in Western Skagerrak, $54 \%$ of total landings for the stock in 2015) is slightly higher than in 2014.
Since 2012, the fishing pattern has moved slightly away from the North Sea boundary rectangle (43F8) towards the more central Skagerrak (44F9) (Figure 7.4.2).

Plaice in Skagerrak is fished in the relatively shallow waters of the Western Skagerrak $(<50 \mathrm{~m})$, where it is the main target species. Therefore, CPUE trends were also computed for the small vessels separately ( $<10 \mathrm{~m}$, without VMS) as these fish closer to the shore, and the trends were also pointing upwards for this segment. (Figure 7.4.3).

### 7.3.3 Survey series

The benchmark WKPLE recommended to remove the haul with high CPUE from the survey index, until enough additional coastal hauls are taken to account for the effect of depth on the index, for example using Berg et al. GAM model.

Trends in IBTS are different between spring and autumn (Figure 7.4.5). The autumn survey seems to show consistent high signals for some year classes, and in particular at age 3 the picks correspond to the large year classes 2007, 2004 and 2002 observed in the North Sea. Based on the additional analyses of seasonal patterns performed during WKPLE, this confirms the hypothesis of summer inflow from the North Sea into the Skagerrak, indicating that Skagerrak belongs to the distribution area of North Sea plaice. The spring survey is less consistent, and does not track populations over time very well (Figure 7.4.6). The improper IBTS sampling design for plaice might explain
part of the issue, but it is also hypothesised that spawning in Skagerrak is not very regular, and that the densities of populations may vary from year to year. This is also supported by the hydrodynamic simulations presented in WKPLE, which demonstrated large variability from year to year of the inflow of North Sea water into the Skagerrak. Nevertheless, there is a significant correlation between the estimated recruitment of the combined stock and the observed recruitment in Skagerrak in IBTS Q1 age 1 (Figure 7.4.7).

### 7.4 Assessment

The analytical assessment of the combined stock of plaice in the North Sea and Skagerrak is presented in section 8.

### 7.5 References

ICES. 2015. Report of the Benchmark Workshop on Plaice (WKPLE), 23-27 February 2015, ICES Headquarters, Copenhagen, Denmark. ICES CM 2015 \ACOM:33. 200 pp.
Ulrich, C., Boje, J., Cardinale, M., Gatti, P., LeBras, Q., Andersen, M., Hemmer-Hansen, J., Hintzen, N.T., Jacobsen, J.B., Jonsson, P., Miller, D.C.M., Nielsen, E.E., Rijnsdorp, A.D., Sköld, M., Svedäng, H., Wennhage, H. 2013. Variability and connectivity of plaice populations from the Eastern North Sea to the Western Baltic Sea, and implications for assessment and management, Journal of Sea Research, 10.1016/j.seares.2013.04.007.
Ulrich, C., Hemmer-Hansen, J., Boje, J., Christensen, A., Hüssy, K., Clausen, L.A.W., in press, Variability and connectivity of plaice populations from the Eastern North Sea to the Baltic Sea, Part II. Biological evidence of population mixing. Journal of Sea Research..

Table 7.1.1 Plaice in Subdivision 20 (Skagerrak). ICES estimates of landings by country in tonnes.

| Year | Denmark | Sweden | Germany | Belgium | Norway | Netherlands | Total | DISC <br> RAT | TAC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 5095 | 70 |  |  | 3 |  | 5168 |  |  |
| 1973 | 3871 | 80 |  |  | 6 |  | 3957 |  |  |
| 1974 | 3429 | 70 |  |  | 5 |  | 3504 |  |  |
| 1975 | 4888 | 77 |  |  | 6 |  | 4971 |  |  |
| 1976 | 9251 | 51 |  | 717 | 6 |  | 10025 |  |  |
| 1977 | 12855 | 142 |  | 846 | 6 |  | 13849 |  |  |
| 1978 | 13383 | 94 |  | 371 | 9 |  | 13857 |  |  |
| 1979 | 11045 | 67 |  | 763 | 9 |  | 11884 |  |  |
| 1980 | 9514 | 71 |  | 914 | 11 |  | 10510 |  |  |
| 1981 | 8115 | 110 |  | 263 | 13 |  | 8501 |  |  |
| 1982 | 7789 | 146 |  | 127 | 11 |  | 8073 |  |  |
| 1983 | 6828 | 155 |  | 133 | 14 |  | 7130 |  |  |
| 1984 | 7560 | 311 |  | 27 | 22 |  | 7920 |  |  |
| 1985 | 9646 | 296 |  | 136 | 18 |  | 10096 |  |  |
| 1986 | 10645 | 202 |  | 505 | 26 |  | 11378 |  |  |
| 1987 | 11327 | 241 |  | 907 | 27 |  | 12502 |  |  |
| 1988 | 9782 | 281 |  | 716 | 41 |  | 10820 |  |  |
| 1989 | 5414 | 320 |  | 230 | 33 |  | 5997 |  |  |
| 1990 | 8729 | 779 |  | 471 | 69 |  | 10048 |  |  |
| 1991 | 5809 | 472 | 15 | 315 | 68 |  | 6679 |  |  |
| 1992 | 8514 | 381 | 16 | 537 | 106 |  | 9554 |  | 11200 |
| 1993 | 9125 | 287 | 37 | 326 | 79 |  | 9854 |  | 11200 |
| 1994 | 8783 | 315 | 37 | 325 | 91 |  | 9551 |  | 11200 |
| 1995 | 8468 | 337 | 48 | 302 | 224 |  | 9379 |  | 11200 |
| 1996 | 7304 | 260 | 11 |  | 428 |  | 8003 |  | 11200 |
| 1997 | 7306 | 244 | 14 |  | 249 |  | 7813 |  | 11200 |
| 1998 | 6132 | 208 | 11 |  | 98 |  | 6449 |  | 11200 |
| 1999 | 6473 | 233 | 7 |  | 336 |  | 7049 |  | 11200 |
| 2000 | 6680 | 230 | 5 |  | 67 |  | 6982 |  | 11200 |
| 2001 | 9045 | 125 |  |  | 61 |  | 9231 |  | 9400 |
| 2002 | 6773 | 141 | 3 |  | 164 | 3 | 7084 |  | 6400 |
| 2003 | 5079 | 143 | 8 |  | 385 | 1484 | 7098 |  | 10400 |
| 2004 | 5999 | 545 | 67 |  | 111 | 1288 | 8011 |  | 9500 |
| 2005 | 4684 | 554 | 14 |  | 9 | 823 | 6084 |  | 7600 |
| 2006 | 6563 | 366 | 21 |  | 352 | 1059 | 8361 |  | 7600 |
| 2007 | 5656 | 281 | 21 |  | 166 | 1503 | 7626 |  | 8500 |
| 2008 | 7163 | 220 | 17 |  | 117 | 775 | 8292 |  | 9300 |
| 2009 | 5828 | 92 | 13 |  | 62 | 506 | 6500 |  | 9300 |
| 2010 | 7101 | 127 | 13 |  | 103 | 1331 | 8676 |  | 9300 |
| 2011 | 7746 | 179 | 13 |  | 230 | 15 | 8183 |  | 7900 |
| 2012 | 7338 | 155 | 12 |  | 136 | 10 | 7651 | 12\% | 7900 |
| 2013 | 6326 | 160 | 10 |  | 138 | 181 | 6815 | 14\% | 9142 |
| 2014 | 7484 | 240 | 46 |  | 48 | 506 | 8981 | 10\% | 10056 |
| 2015 | 7808 | 274 | 14 |  | 69 | 1639 | 9804 |  | 10056 |
| 2016 |  |  |  |  |  |  |  |  | 11766 |

Table 7.3.3 Summary of data provided to InterCatch 2002-2015

|  | previous landings | revised landings | revision | discards provided to IC | total discards after raising | discards <br> ratio | share of landings with discards provided | share of landings with age information provided |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | t | t | \% | t | t | \% |  |  |
| 2002 | 6671 | 7087 | 6\% | 517 | 574 | 7\% | 0.89 | 0.96 |
| 2003 | 6656 | 7100 | 7\% | 748 | 1437 | 17\% | 0.51 | 0.72 |
| 2004 | 7513 | 8013 | 7\% | 1761 | 2873 | 26\% | 0.59 | 0.75 |
| 2005 | 5690 | 6084 | 7\% | 1200 | 2081 | 25\% | 0.62 | 0.77 |
| 2006 | 7855 | 8360 | 6\% | 1309 | 2243 | 21\% | 0.53 | 0.78 |
| 2007 | 7406 | 7626 | 3\% | 1714 | 2862 | 27\% | 0.55 | 0.74 |
| 2008 | 7607 | 8295 | 9\% | 811 | 1043 | 11\% | 0.72 | 0.86 |
| 2009 | 6035 | 6502 | 8\% | 520 | 610 | 9\% | 0.87 | 0.9 |
| 2010 | 9187 | 8676 | -6\% | 661 | 842 | 9\% | 0.81 | 0.82 |
| 2011 | 8342 | 8183 | -2\% | 919 | 1040 | 11\% | 0.94 | 0.95 |
| 2012 | 7627 | 7651 | 0\% | 734 | 846 | 10\% | 0.86 | 0.96 |
| 2013 | 6825 | 6815 | 0\% | 949 | 1161 | 15\% | 0.81 | 0.93 |
| 2014 |  | 8981 |  | 836 | 1022 | 10\% | 0.8 | 0.83 |
| 2015 |  | 9804 |  | 524 | 677 | 6\% | 0.78 |  |

Table 7.3.4. Plaice in Skagerrak. Landings number at age.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 1 | 809 | 8059 | 9177 | 3915 | 1760 | 375 | 73 | 25 | 23 |
| 1985 | 1 | 142 | 3816 | 17915 | 5815 | 1633 | 624 | 154 | 116 | 97 |
| 1986 | 1 | 3 | 2172 | 12185 | 17220 | 3886 | 509 | 214 | 107 | 152 |
| 1987 | 1 | 16 | 1814 | 8845 | 16315 | 9804 | 1983 | 293 | 167 | 121 |
| 1988 | 1 | 33 | 1922 | 10081 | 12460 | 6358 | 2512 | 803 | 254 | 148 |
| 1989 | 1 | 296 | 2256 | 6024 | 5530 | 2404 | 1032 | 468 | 194 | 216 |
| 1990 | 1 | 1311 | 6462 | 7785 | 9284 | 3084 | 888 | 436 | 319 | 358 |
| 1991 | 1 | 851 | 5312 | 8195 | 4480 | 2810 | 828 | 268 | 129 | 162 |
| 1992 | 1 | 54 | 1406 | 9159 | 16174 | 4146 | 932 | 260 | 89 | 71 |
| 1993 | 1 | 224 | 2369 | 9351 | 12579 | 6392 | 1381 | 309 | 82 | 43 |
| 1994 | 1 | 19 | 5087 | 7295 | 9521 | 7596 | 2129 | 292 | 91 | 34 |
| 1995 | 1 | 0 | 655 | 5404 | 11006 | 6475 | 4848 | 843 | 119 | 69 |
| 1996 | 1 | 863 | 3517 | 6322 | 4849 | 4609 | 1768 | 1318 | 137 | 25 |
| 1997 | 1 | 0 | 541 | 4647 | 8783 | 4875 | 2985 | 1332 | 832 | 121 |
| 1998 | 1 | 198 | 4783 | 5307 | 5991 | 2700 | 685 | 348 | 210 | 200 |
| 1999 | 1 | 0 | 1160 | 6174 | 7456 | 7234 | 1239 | 361 | 71 | 129 |
| 2000 | 1 | 0 | 1114 | 7270 | 10566 | 3276 | 854 | 109 | 10 | 22 |
| 2001 | 1 | 1035 | 5422 | 8212 | 10722 | 4540 | 288 | 76 | 8 | 33 |
| 2002 | 0 | 70 | 1642 | 6928 | 7508 | 5106 | 1848 | 458 | 49 | 38 |
| 2003 | 11 | 2497 | 3005 | 6189 | 7784 | 4396 | 1057 | 145 | 19 | 12 |
| 2004 | 0 | 1661 | 16204 | 3693 | 3105 | 1930 | 320 | 133 | 13 | 16 |
| 2005 | 34 | 2330 | 3707 | 9036 | 3186 | 1401 | 597 | 145 | 64 | 12 |
| 2006 | 0 | 770 | 10525 | 7991 | 7264 | 1304 | 480 | 185 | 46 | 16 |
| 2007 | 146 | 2080 | 4306 | 8357 | 5113 | 4441 | 569 | 133 | 53 | 14 |
| 2008 | 18 | 2109 | 8966 | 6659 | 4224 | 1736 | 1170 | 60 | 55 | 28 |
| 2009 | 0 | 1074 | 8310 | 6529 | 2654 | 1085 | 217 | 90 | 1 | 6 |
| 2010 | 78 | 1795 | 8819 | 10703 | 3102 | 735 | 261 | 81 | 62 | 14 |
| 2011 | 183 | 3022 | 4299 | 6794 | 5139 | 1770 | 310 | 166 | 65 | 20 |
| 2012 | 162 | 1668 | 8012 | 5105 | 2820 | 1696 | 248 | 157 | 101 | 89 |
| 2013 | 0 | 2753 | 9401 | 4436 | 1102 | 777 | 365 | 94 | 49 | 42 |
| 2014 | 0 | 653 | 11091 | 9280 | 3522 | 895 | 481 | 245 | 64 | 50 |
| 2015 | 0 | 1406 | 9947 | 11611 | 3853 | 1059 | 223 | 170 | 58 | 23 |

Table 7.2.2. Plaice in Skagerrak. Landings weight at age.

| CW | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 |  | 0.276 | 0.299 | 0.301 | 0.373 | 0.423 | 0.548 | 0.817 | 1.029 | 1.319 |
| 1985 |  | 0.212 | 0.294 | 0.309 | 0.351 | 0.434 | 0.55 | 0.759 | 0.872 | 0.993 |
| 1986 |  | 0.395 | 0.26 | 0.28 | 0.304 | 0.379 | 0.543 | 0.736 | 0.94 | 1.041 |
| 1987 |  | 0.205 | 0.245 | 0.266 | 0.285 | 0.358 | 0.525 | 0.728 | 0.911 | 1.127 |
| 1988 |  | 0.22 | 0.251 | 0.261 | 0.285 | 0.343 | 0.466 | 0.551 | 0.746 | 1.111 |
| 1989 |  | 0.216 | 0.24 | 0.274 | 0.315 | 0.372 | 0.465 | 0.639 | 0.703 | 0.876 |
| 1990 |  | 0.267 | 0.28 | 0.289 | 0.333 | 0.389 | 0.484 | 0.667 | 0.756 | 1.077 |
| 1991 |  | 0.27 | 0.26 | 0.248 | 0.27 | 0.361 | 0.49 | 0.577 | 0.653 | 1.032 |
| 1992 |  | 0.274 | 0.318 | 0.265 | 0.278 | 0.334 | 0.506 | 0.67 | 0.85 | 0.872 |
| 1993 |  | 0.229 | 0.25 | 0.266 | 0.291 | 0.338 | 0.456 | 0.581 | 0.669 | 0.884 |
| 1994 |  | 0.365 | 0.246 | 0.265 | 0.286 | 0.33 | 0.41 | 0.586 | 0.653 | 0.785 |
| 1995 |  |  | 0.297 | 0.296 | 0.286 | 0.325 | 0.366 | 0.498 | 0.726 | 0.767 |
| 1996 |  | 0.225 | 0.252 | 0.282 | 0.384 | 0.399 | 0.437 | 0.428 | 0.559 | 1.013 |
| 1997 |  |  | 0.248 | 0.266 | 0.291 | 0.335 | 0.408 | 0.458 | 0.441 | 0.492 |
| 1998 |  | 0.226 | 0.242 | 0.273 | 0.328 | 0.401 | 0.468 | 0.513 | 0.574 | 0.655 |
| 1999 |  |  | 0.277 | 0.294 | 0.287 | 0.292 | 0.33 | 0.357 | 0.661 | 0.585 |
| 2000 |  |  | 0.24 | 0.273 | 0.301 | 0.351 | 0.38 | 0.489 | 0.857 | 0.911 |
| 2001 |  | 0.257 | 0.282 | 0.292 | 0.322 | 0.306 | 0.423 | 0.604 | 0.876 | 0.658 |
| 2002 | 0 | 0.221 | 0.285 | 0.276 | 0.277 | 0.317 | 0.39 | 0.459 | 0.751 | 1.064 |
| 2003 | 0.217 | 0.239 | 0.258 | 0.272 | 0.292 | 0.296 | 0.401 | 0.424 | 0.669 | 1.064 |
| 2004 | 0 | 0.238 | 0.269 | 0.324 | 0.348 | 0.381 | 0.481 | 0.654 | 0.791 | 0.97 |
| 2005 | 0.227 | 0.251 | 0.259 | 0.292 | 0.325 | 0.359 | 0.397 | 0.529 | 0.659 | 1.087 |
| 2006 | 0 | 0.255 | 0.264 | 0.287 | 0.326 | 0.32 | 0.343 | 0.424 | 0.686 | 1.063 |
| 2007 | 0.246 | 0.249 | 0.271 | 0.304 | 0.309 | 0.334 | 0.34 | 0.47 | 0.476 | 1.073 |
| 2008 | 0.239 | 0.282 | 0.314 | 0.315 | 0.356 | 0.412 | 0.466 | 0.558 | 0.579 | 0.533 |
| 2009 | 0 | 0.241 | 0.281 | 0.332 | 0.393 | 0.488 | 0.529 | 0.658 | 1.121 | 0.872 |
| 2010 | 0.102 | 0.247 | 0.31 | 0.334 | 0.404 | 0.538 | 0.681 | 0.693 | 0.677 | 1.08 |
| 2011 | 0.258 | 0.302 | 0.322 | 0.354 | 0.431 | 0.484 | 0.679 | 0.733 | 0.585 | 1.356 |
| 2012 | 0.236 | 0.276 | 0.347 | 0.368 | 0.437 | 0.519 | 0.606 | 0.6 | 0.828 | 0.916 |
| 2013 | 0 | 0.287 | 0.319 | 0.376 | 0.47 | 0.595 | 0.631 | 0.767 | 0.778 | 0.825 |
| 2014 | 0 | 0.274 | 0.301 | 0.352 | 0.406 | 0.486 | 0.634 | 0.676 | 0.757 | 0.902 |
| 2015 | 0 | 0.289 | 0.326 | 0.354 | 0.401 | 0.473 | 0.639 | 0.753 | 0.715 | 0.961 |

Table 7.2.1. Plaice in Skagerrak. Discards number at age

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 620 | 828 | 2050 | 1827 | 1263 | 57 | 2 | 0 | 0 | 0 |
| 2003 | 80 | 5695 | 3998 | 1054 | 515 | 33 | 5 | 0 | 0 | 0 |
| 2004 | 746 | 3947 | 12130 | 4036 | 928 | 218 | 4 | 0 | 0 | 0 |
| 2005 | 351 | 7413 | 3157 | 4323 | 396 | 39 | 140 | 5 | 0 | 0 |
| 2006 | 481 | 3094 | 10099 | 652 | 44 | 2 | 4 | 1 | 0 | 0 |
| 2007 | 448 | 6579 | 6429 | 4911 | 858 | 56 | 20 | 0 | 0 | 0 |
| 2008 | 165 | 4484 | 2923 | 745 | 132 | 6 | 7 | 0 | 0 | 0 |
| 2009 | 27 | 1105 | 2754 | 299 | 33 | 2 | 2 | 0 | 0 | 0 |
| 2010 | 1473 | 2218 | 1458 | 131 | 23 | 10 | 1 | 2 | 1 | 0 |
| 2011 | 606 | 3291 | 1919 | 529 | 158 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 823 | 2850 | 1717 | 360 | 31 | 12 | 0 | 0 | 0 | 0 |
| 2013 | 511 | 5531 | 2563 | 690 | 21 | 6 | 0 | 0 | 0 | 0 |
| 2014 | 474 | 3632 | 3741 | 831 | 119 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 66 | 1050 | 2134 | 996 | 120 | 35 | 14 | 1 | 0 | 0 |

Table 7.2.2. Plaice in Skagerrak. discard weight at age.

| CW | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 0.03 | 0.063 | 0.089 | 0.094 | 0.115 | 0.058 | 0.176 | 0.218 | 0.151 | 0 |
|  | 003 | 0.055 | 0.105 | 0.149 | 0.143 | 0.156 | 0.119 | 0.15 | 0.237 | 0 |
| 2004 | 0.066 | 0.097 | 0.134 | 0.155 | 0.169 | 0.133 | 0.118 | 0 | 0 | 0 |
| 2005 | 0.075 | 0.134 | 0.139 | 0.155 | 0.153 | 0.119 | 0.187 | 0.11 | 0 | 0 |
| 2006 | 0.047 | 0.115 | 0.175 | 0.142 | 0.157 | 0.251 | 0.186 | 0.223 | 0 | 0 |
| 2007 | 0.051 | 0.125 | 0.155 | 0.176 | 0.172 | 0.142 | 0.17 | 0 | 0 | 0 |
| 2008 | 0.047 | 0.105 | 0.134 | 0.174 | 0.166 | 0.148 | 0.445 | 0 | 0 | 0 |
| 2009 | 0.032 | 0.126 | 0.151 | 0.162 | 0.169 | 0.139 | 0.128 | 0 | 0 | 0 |
| 2010 | 0.098 | 0.158 | 0.195 | 0.258 | 0.393 | 0.385 | 0.8 | 0.618 | 0.474 | 0 |
| 2011 | 0.084 | 0.156 | 0.175 | 0.19 | 0.235 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0.083 | 0.145 | 0.169 | 0.183 | 0.212 | 0.209 | 0.6 | 0 | 0.633 | 0 |
| 2013 | 0.071 | 0.113 | 0.147 | 0.172 | 0.203 | 0.23 | 0 | 0 | 0 | 0 |
| 2014 | 0.049 | 0.076 | 0.151 | 0.165 | 0.174 | 0 | 0 | 0.418 | 0 | 0 |
| 2015 | 0.043 | 0.11 | 0.16 | 0.179 | 0.217 | 0.248 | 0.189 | 0.562 | 0 | 0 |




Figure 7.1.1. Plaice Skagerrak. Upper: total landings 1974-2015. Lower: Landings vs. TAC in Skagerrak, 1992-2016.



Figure 7.3.1. Landings strata with discards ratios available (top) and with age information available (down) as provided to InterCatch before raising.



Figure 7.3.2. Landings at age and mean weight at age in the landings.


Figure 7.3.3. Discards number in Skagerrak



Figure 7.4.1 Trends in EU Landings, effort and lpue by gear and Skagerrak area. G=Gillnets , $1=$ TR1, 2= TR2.(STECF data)


Figure 7.4.2. EU plaice Landings 2006-2015 (from left to right then top to bottom) for the main fisheries (TR1, TR2, GN1) all gears (STECF data). Bubble max size= $\mathbf{3 2 0 0}$ tonnes.


Figure 7.4.3. Landings and lpue by quarter and fleet segment showing both target fishery (TR1) and bycatch fishery (TR2), for small inshore vessels ( $<10 \mathrm{~m}$ ) and larger vessels ( $>10 \mathrm{~m}$ ). STECF data. 2014 provisional.


Figure 7.4.5. Standardized mean CPUE by haul, IBTS Q1 and Q3, for Skagerrak only and for the combined area North Sea plus Skagerrak.



Figure 7.4.7 Correlation between IBTS Q1 at age 1 and the estimated recruitment of the combined stock.

## 8 Plaice (Pleuronectes platessa) in Subarea 4 (North Sea) and Division 3.a (Skagerrak)

A Stock Annex is available for North Sea plaice. Therefore only a comprehensive description of the methods and deviations from the stock annex are presented within this Section of the report. In 2015, the stock annex was updated and two significant changes took place: 1) the North Sea stock is now assessed in a combined assessment with the Skagerrak and 2) the SNS survey is now split into two surveys, from 1982-1999 and from 2000-present (2015).

### 8.1 General

### 8.1.1 Stock structure

The flatfish benchmark group (WKFLAT, ICES, 2010) recommended to explore the potential to perform an integrated assessment of the continuum of plaice stocks from the Baltic to the English Channel. ICES evaluated the stock identity of plaice in the Skagerrak and Kattegat during a dedicated workshop (WKPESTO; ICES 2012b) for which until now combined advice was given.

Plaice in the Skagerrak is considered to have two components: an Eastern and Western. The latter occurs in a mix with plaice migrating in from the North Sea (Ulrich et al 2013) and the predominance of catches occurs on summer feeding aggregations in the Western Skagerrak. In a benchmark (WKPLE 2015, ICES 2015) it was decided that plaice in the Skagerrak would be assessed together with the North Sea stock.

In addition, as in previous years, part of the catches in the 7.d area in the first quarter are included in the North Sea plaice assessment, since North Sea plaice migrates into the area in that season (ICES 2010). This year, $50 \%$ of the mature animals from 7.d in Q1 were added to the North Sea stock, whereas before, $50 \%$ of the total catches were added. Moreover, this year $50 \%$ of the mature discards in Q1 were also added to the North Sea stock.

### 8.1.2 Ecosystem considerations

Available information on ecosystem aspects can be found in the Stock Annex. In addition, the ICES Working Group on the Ecosystem Effects of Fishing Activities (WGECO, ICES 2014b) met in April 2014 and addressed a specific question in relation to North Sea plaice, in response to a request from WGNSSK in 2013:
"According to WGNSSK estimates, the North Sea is currently ongoing a plaice outburst without precedent. However, plaice is not included in multispecies models, so the consequences of this outburst on the North Sea ecosystem are unclear and would potentially require additional focus".
WGECO addressed the trends shown in the stock assessment of plaice, which show how increasing fishing pressure on the stock has progressively moved SSB away from the desired state (in the 1980s and 1990s), and then how management has rectified this situation in recent years, which has brought the North Sea plaice stock in a situation unlike any other over the whole 58 year period for which data is available. The group investigated a possible relationship of these trends with abundance of benthic biomass, which is a predominant food source for plaice. Q1 IBTS data showed a two-fold increase in demersal benthivore biomass over the last 29 year period of the survey, and
that species composition of the demersal benthivore guild has changed as well. The data showed that predation loading by plaice on benthic invertebrates increased by a factor of 8.8 in just eleven years (2000-2011).

The increase in the consumption of benthic invertebrate prey by the whole demersal benthivore guild, and particularly by plaice, raises the question as to whether the abundance of benthic invertebrate prey might be becoming limiting. If the biomass of demersal benthivorous fish is approaching its carrying capacity, then growth rates in the dominant species in the guild might start to decline (which is in this case plaice growth rates). Computed growth coefficients for the 1956 to 2002 cohorts showed a strong declining linear trend over the whole period (albeit with clear systematic variation in the residuals), and this has been related to increasing water temperature in the North Sea. However, fitting a 4th order polynomial function to the data suggested a marked decline in cohort growth towards the end of the time-series. This is perhaps indicative of plaice becoming food limited, possibly suggesting that $B_{M S Y}$ targets for the stock might be marginally too high to be supported by available benthic invertebrate food supplies. However, this evidence is by no means conclusive as polynomial functions are known to show a tendency for marked swings at the extremes of the data range. The situation will become clearer in a few years' time when data for more recent cohorts can be added to the analysis.

On another issue, moving towards better informed estimations of natural mortality (M) may be timely, since fishing mortality now is at a relatively low level, and the assessment model will become more sensitive to the assumption of M. Plaice is not usually included in multi-species models, since it is commonly assumed that plaice is not a common prey species. Another source of natural mortality may be disease outbreaks. Such outbreaks may be more severe when population density and thus population size is high. McVicar and McLay (1985) showed that plaice are relatively susceptible to Ichthyophonus. Patterson (1996) showed that an outbreak of this fungus was responsible for the mortality of approximately $10 \%$ of the North Sea herring population between 1991 and 1994. In order to obtain more realistic estimates of $M$ for plaice in the near future it may be worthwhile to consider the impact of diseases.

### 8.1.3 Fisheries

A basic description of the fisheries is available in the Stock Annex. In recent years, the adoption of innovative gears - which are often aimed at reduction of fuel consumption and reduction of bottom disturbance - may be contributing to changes in fishing patterns however. In 2011, approximately 30 derogation licenses for Pulse trawls were taken into operation, which increased to 42 in 2012. An additional 42 derogation licenses have been extended in spring 2014. At the same time, possible amendments to EU regulations which would permanently legalize the use of pulse gears for the whole fleet are ongoing. Potential future impact either on the plaice stock itself or the stock assessment is unknown. ICES recommends that further studies aimed at investigating catch composition of these innovative gears in comparison to traditional beam trawls are undertaken.

### 8.1.4 ICES Advice

The information in this section is taken from the ICES advice sheet 2015, section 6.3.31:
ICES advises that when the second stage of the EU management plan (Council Regulation No. 676/2007) is applied, catches in 2016 should be no more than 213440 tonnes in Subarea 4 and Division 3.a (Skagerrak part) combined. If this stock is not under the

EU landing obligation in 2016 and discard rates do not change from the average (20122014), this implies landings of no more than 159197 tonnes.

## Management plan

The plan consists of two stages and is now in stage two; implementation of this second stage (as stipulated in article 5 of the EC regulation) is not yet defined. Application of stage two of the plan is based on transitional arrangements until an evaluation of the plan has been conducted (as stipulated in article 5 of the EC regulation). ICES is using the existing management plan for advice on the combined stock.

### 8.1.5 Management

A multiannual plan for plaice and sole in the North Sea was adopted by the EU Council in 2007 (EC regulation 676/2007) describing two stages of which the first stage should be deemed a recovery plan and its second stage a management plan. ICES has evaluated the plan as in agreement with the precautionary approach (Miller and Poos 2010; Simmonds 2010). A subsequent evaluation in 2012 (Coers et al, 2012) addressed amendments to the plan in the context of moving towards stage two of the plan. These amendments do not affect the current advice for plaice.

### 8.2 Data available

### 8.2.1 Landings

During the benchmark of the eastern channel (7.d) plaice stock (WKFLAT) it was decided that $50 \%$ of Q1 mature fish catches taken in the eastern channel are actually plaice from the North Sea stock migrating in and out of the area. Before 2015, 50\% of the Q1 eastern channel (7.d) plaice landings were included in the assessment of the North Sea plaice stock. Since 2015, $50 \%$ of the mature fish in the landings in Q1 and of the mature fish in the discards in Q1 were added to the North Sea stock and the time series was updated, such that in previous years also $50 \%$ of the mature catches from Q1 were added. See the stock annex for plaice in division 7.d for further details.

During the benchmark on plaice (WKPLE, ICES 2015) it was decided that plaice from the Skagerrak would be added to the North Sea stock. Since, the assessment is a combined assessment with Skagerrak plaice.

Total landings (including 7.d and Skagerrak) of North Sea plaice in 2015 were estimated by the WG at 85360 t . Of these 74963 t came from the North Sea (excluding Skagerrak). This is an increase of $6 \%$ from the 2014 landings and only $58 \%$ of the 128376 t TAC for 2015. Total landings (in tonnes) are presented in Table 8.2.1 and landings in numbers at age in Table 8.2.2.

### 8.2.2 Discards

The discards time series used in the assessment includes Dutch, Danish, German and UK discards observations for 2000-2015, as is described in the stock annex. From Belgium, discards data have been available as well but were only used in the assessment since 2012, since it became available through InterCatch. See section 8.2 .7 for more information on the use of InterCatch for raising discards rates across metiers and countries. The Dutch discards data for 2009 and 2010 were derived from a combination of the observer programme that has been running since 2000, and a new self-sampling programme. The estimates from both programmes were combined to come up with an overall estimate of discarding by the Dutch beam trawl fleet. Since 2011, estimates were
derived exclusively from the self-sampling data. There is an on-going project within IMARES to validate these estimates by examining matched (same vessel and haul) trips where both observer estimates and self-sampling estimates are derived.
To reconstruct the number of plaice discards at age before 2000, catch numbers at age data was reconstructed in 2005 based on a model-based analysis of growth, selectivity of the $80-\mathrm{mm}$ beam trawl gear, and the availability of undersized plaice on the fishing grounds. Discards numbers at age are presented in Table 8.2.3. Figure 8.2.1 presents a time series of landings, catches and discards from these different sources.

### 8.2.3 Catch

The total catch at age as used in the assessment including all landings and all discards are presented in Table 8.2.4. These include catch of NS plaice in the $1^{\text {st }}$ quarter from division 7.d and catch from the Skagerrak. Landings-at-age, discards-at-age and catch-at-age plots are presented in figures 8.2.2 and 8.2.3.

### 8.2.4 Weight at age

Stock weights at age are presented in Table 8.2.5. Stock weight at age has varied considerably over time, especially for the older ages. Landing, discards and catch weights at age are presented in Table 8.2.6, 8.2.7 and 8.2.8 respectively. Catch weights at age are derived from the discards and landings weights at age according to the relative contributions of each to the overall catch for each age. Figure 8.2.4 presents the stock, discards, landings and catch weights at age. Notably, there has been a long-term decline in the observed stock weight at age.

### 8.2.5 Maturity and natural mortality

Natural mortality is assumed to be 0.1 for all age groups and constant over time. A fixed maturity ogive (Table 8.2.9) is used for the estimation of SSB in North Sea plaice.

### 8.2.6 Catch, effort and survey data

Three survey indices are used as tuning indices, as decided during an Inter Benchmark Protocol, in March 2013 (Miller and Coers 2013). For some additional explanation, see also the WGNSSK report of 2013 (ICES 2013). This year, the SNS survey was split into two timeseries.

Table 8.2.10 and Figure 8.2.5 show the index values for the years that they are used in the assessment:

Beam Trawl Survey combined for RV Tridens and ISIS (BTS-combined); (1996-2015)
Beam Trawl Survey RV Isis (BTS-Isis) for the older part of the time series; (1985-1995)
Sole Net Survey 1 (SNS1); (1982-1999)
Sole Net Survey 2 (SNS2); (2000-2015)
Of the BTS-combined survey index, ages 1-9 are used for tuning the North Sea plaice assessment. Of the BTS-Isis older survey index, ages 1-8 are used. And of the Sole Net Survey (SNS1 \& SNS2) ages 1-3 are used in the assessment, while the 0-group index is used in the RCT3 analysis for recent recruitment estimates. The internal consistency of the survey indices used for tuning appears relatively high for the Beam trawl surveys, but low for the SNS surveys (Figures 8.2.6-8.2.8).

Since 2011 there is an annual survey of plaice and sole using commercial vessels and gears (Reijden et al. 2016). This survey takes place in the same season as the BTS surveys. Length structured catch per unit effort estimates and age-length keys are collected during this survey.
An additional survey index is used for recruitment estimates in the RCT3 analysis (Table 8.5.1):
Demersal Fish Survey (DFS) age-0
Several commercial LPUE series consisting of an effort series and landings-at-age series are available for usage as tuning fleets. These include time series for the Dutch beam trawl fleet and the UK beam trawl fleet (excluding all flag vessels). Because WKFLAT 2009 recommended to exclude LPUE series from the final assessment run upon which management advice is based, they have not been included in the assessment.

### 8.2.7 InterCatch

Since 2012, national research institutes submitted landings and discard estimates by métier and quarter in InterCatch. Figure 8.2.9 shows the landings and discards by country and by métier in area 4 . Approximately $54 \%$ of the landings in weight are sampled to obtain information on age-composition (Note that the UK vessels of the TBB_DEF_70-99_mm metier are exclusively Dutch owned flag vessels and de facto are thus sampled in the Dutch market sampling programme). Of the metiers for which discards are monitored in sampling programmes, the largest part of these discards is covered in the TBB_DEF_70-99_mm fleet. In most discards monitoring programmes, age composition information is also collected. Approximately $94 \%$ of the discards (in weight) were sampled. To raise the amount of discards for landings that had no discards allocated and to raise the landings and discards for which no age distribution was known, the same following allocation scheme was used. Allocations to calculate the age structure were done separately for discards and landings. The métiers that covered most of the catches each had their own group (OTB 70-119, OTB > 120, TBB 70119, TBB > 120 and OTB \& TBB CRU, see table below). Other countries that had sampled the métiers were used to allocate discard and age structure to the unsampled fleets. All other métiers were grouped into one group. All métiers except the métiers for crustaceans (_CRU) were used to allocate discards and age structure to this group. All allocations were done per quarter. If age structures were present for data for the whole year only, these were added to all quarters. If there were no samples in a specific quarter, all other quarters were used. No discards were sampled for TBB $>120$, therefore OTB > 120 was used for this group.
Allocation scheme to raise discards and age structures to unsampled fleets.

| UNSAMPLED FLEET* | SAMPLED FLEET** |
| :--- | :--- |
| OTB 70-119 | OTB 70-119 |
| OTB > 120 | OTB > 120 |
| TBB 70-119 | TBB 70-119 |
| TBB > 120 | TBB > 120 ( OTB > 120) |
| OTB \& TBB CRU | OTB \& TBB CRU |
| Others | All métiers, excluding métiers for crustaceans (_CRU) |

* Unsampled fleet are those fleets for which no dicards or age structure is known.
${ }^{* *}$ Sampled fleet are those fleets for which the discard rate or age structure is known.


### 8.2.8 Data analyses

The assessment of North Sea plaice by XSA was carried out using the FLR (FLCore v. 2.3 and FLXSA v.2.0) in R version 2.13. All other post-analyses were done using FLR packages.
Since 2013, ICES does not operate with external review groups anymore. Audits were done by internal reviewers (members of the WGNSSK group) and potential issues were directly discussed between the auditors and the stock assessor. Therefore there is no written review to be presented here.

### 8.3 Assessment

### 8.3.1 Exploratory catch-at-age-based analyses

Additional exploratory assessments were run. Since 2011, there is a survey using commercial vessels held annually in the Netherlands (van der Reijden et al. 2016). The indices from this survey were used in an exploratory assessment, combined with the 4 tuning indices that are currently used. The index values per age are given in Figure 8.3.1. The results for the two assessments are compared in Figure 8.3.2. Although the two assessments are very much alike, the survey with the commercial gears results in lower SSB estimates in the most recent 10 years, and slightly higher F values.

## Conclusions exploratory runs

The group agreed that incorporation of the survey with the commercial gears and vessels should be considered at the next benchmark.

## Final XSA assessment

The settings for the final assessment that is used for the catch option table is given below:

| Stock | North Sea and SKaGerrak combined |
| :--- | :--- |
| Year | 2016 |
| Catch at age | Landings + (reconstructed) discards based on |
|  | NL, DK + UK + DE fleets and BE (since 2012) |
| Fleets (years; ages) | BTS-Isis-early 1985-1995; 1-8 |
|  | BTS-combined 1996-2015; 1-9 |
|  | SNS1 1982-1999; 1-3 |
|  | SNS2 2000-2015 (excl. 2003); 1-3 |
| Plus group | 10 |
| First tuning year | 1982 |
| Last data year | 2015 |
| Time series weights | No taper |
| Catchability dependent on stock size for age $<$ | 1 |
| Catchability independent of ages for ages >= | 6 |
| Survivor estimates shrunk towards the mean F | 5 years / 5 years |
| s.e. of the mean for shrinkage | 2.0 |
| Minimum standard error for population | 0.3 |
| estimates | Not applied |
| Prior weighting |  |

The full diagnostics are presented in Table 8.3.1. The XSA model converged after 39 iterations. The model fits well to the combined BTS tuning indices (small residuals and
no clear patterns), and also to the early SNS1 tuning index. The later SNS2 time series still shows a pattern with more positive residuals in the earlier years and more negative residuals in the later years, apart from the final year (Figure 8.3.3).
Fishing mortality and stock numbers are shown in Tables 8.3.2 and 8.3.3 respectively. Figure 8.3.4 shows retrospective pattern of the final XSA run with respect to SSB, recruitment and F. Retrospective analyses indicate that each year F and recruitment are corrected upwards and SSB is corrected downwards. It is unknown what causes these patterns. The overall reduction in fishing mortality leads to more uncertain assessments because the share of natural mortality in the total mortality increases. Therefore it is advised to study the impact of varying natural mortalities on the assessment.

### 8.3.2 Final XSA results

Table 8.4.1 shows the stock summary table. Figure 8.4.1 presents the trends in landings, mean $F(2-6)$, SSB, and recruitment since 1957. Reported landings gradually increased up to the late 1980s and then rapidly declined until 1995, in line with the decrease in TAC. The landings show a general decline from 1987 onwards, increasing slowly but steadily in recent years, although in 2014 the landings were slightly lower than in 2013. Discards were particularly high in 1997 and 1998 (reconstructed), and in 2001 and 2003 (observed), resulting from strong year classes. Fishing mortality increased until the late 1990s and reached its highest observed level in 1997. Since the early 2000's, fishing mortality has been rapidly decreasing. Since 2007 it has been below the fishing mortality target established in the management plan. It is currently (2015) estimated at 0.17. Over the last five years SSB has been rapidly increasing and is currently (2015) estimated at 754812 kt . Recruitment varies inter-annually around the long term geometric mean of approximately 1 million recruits. It appears to have been lower on average during the 1960's and 1970's, then above average in the 1980's and fluctuating around the average since the 1990's. In recent years, it is higher than average ( 1.6 million average 2011-2015).

The stock dynamics are affected by the occurrence of strong year classes, but increased stock size in recent years is most likely the direct consequence of reduced fishing mortality, given that no exceptionally strong year classes have been observed in recent years.
The predominant age in the landings is currently age-4 (in 2014 as well as in the past decade, see figure 8.2.2). Notably, during the time series, this was only also observed in the 1960's. In contrast, the predominant age in the landings in the 1970's, 1980's and 1990's, was age-3. The age distribution in the landings in recent years furthermore shows more similarity with the 1960's in that age-5 and age-6 fish are relatively abundant in the landings in comparison to the rest of the time series and age- 2 fish are notably underrepresented in the landings. These shifts in age distribution may be explained by the still relatively low exploitation level in the 1960's, which subsequently substantially increased over the next three decades and since the early 2000's has shown a dramatic decline. Changes in spatial distribution of fishing effort and shifts in spatial distribution of the fish may also have affected these changes. The 'lack' of age2 fish in the landings in the 1960's as well as in recent years may be for a number of reasons. When considering the age distribution in the catches age- 2 fish were also lacking in the catches in the 1960's, while this is not the case in 2014. One possible explanation may be the occurrence of high grading (discarding of smaller fish in order to allow for landing higher numbers of large fish for which a higher price may be received or to avoid exhaustion of quota. The latter seems unlikely since the TAC has not been fully utilised in recent years. Another explanation may be that plaice have become mature
at younger ages than in the past since this shift in maturation also leads to mature fish being of a smaller size at age, because growth rate diminishes after maturation. Grift et al. 2003 observed that this may occur due to fisheries-induced genetic change: those fish that are genetically programmed to mature late at large sizes are likely to have been removed from the population before they have had a chance to reproduce and pass on their genes. This could cause age- 2 fish to be discarded more abundantly in recent years because a larger fraction of them being under the minimum size in comparison to the past.

### 8.3.3 The Fishers' North Sea Stock Survey

The Fishers' North Sea Stock Survey (FNSSS) was carried out using a questionnaire circulated to North Sea fishermen in five countries: Belgium, Denmark, England, the Netherlands, and Scotland. Fishermen were asked to record their perceptions of changes in their economic circumstances, as well as in the state of selected fish stocks. No real relationship was apparent between the plaice abundance index derived from the Fishers' North Sea Stock Survey and the ICES estimates of the North Sea plaice spawning stock biomass.

### 8.4 Recruitment estimates

In the short term forecasts, assumptions are made on a number of things (see also section 8.5). One of the more difficult things to predict is the strength of incoming year classes (abundance of ages 0-2) in the assessment year. A number of options are considered as follows:

Age-0: More specifically, the abundance estimate of age-1 fish in the year after the assessment year, i.e. in the TAC-year) needs to be assumed and no data is available from surveys or otherwise. Therefore, the geometric mean of the time series is used.

Age-1: The RCT3 analysis is run which combines DFS and SNS survey data and the assessment results to predict the abundance of age-1. Depending on the indicated predictive strength of the RCT3 model (typically the magnitude of the standard error) the RCT3 estimate is used in the short-term forecasts. Otherwise, the geometric mean is used.

Age-2: The RCT3 analysis is run which combines DFS, BTS and SNS survey data and the assessment results to predict the abundance of age-2. Depending on the indicated predictive strength of the RCT3 model (typically the magnitude of the standard error) the RCT3 estimate is used in the short-term forecasts. Otherwise the XSA survivors estimate is used.

Input to the RCT3 analysis is presented in Table 8.4.1. The results for age-1 and age-2 abundance estimates are presented in Table 8.4.2, and in Table 8.4.3 respectively. For year class 2014 (age 1 in 2015) the values predicted by the DFS-0 and the SNS-0 survey estimates in RCT3 have similar values and are both lower than the VPA mean respectively. The SNS-0 has a low prediction standard error and the DFS-0 has a high standard error. The WG decided that because the DFS-0 and the SNS-0 had similar values, the RCT3 value was used for the short-term forecasts. For year class 2013 (age 2 in 2015), the estimates from BTS 1-group (comparable to the VPA mean) has a relatively low standard error (compared to the other surveys). However, a retrospective analysis for age- 2 survivors shows that the XSA is relatively strong in predicting age-2 survivors (see retrospective plot, Figure 8.11.1). Hence, the WG decided to use the XSA estimate rather than the RCT3 estimate for the 2014 year class. The recruitment estimates from
the different sources are summarized in the text table below. Underlined values were used in the forecast.

| Year class | Age in 2016 | XSA <br> SURVIVORS | RCT3 | $\begin{gathered} \text { GM } 1957- \\ 2013 \end{gathered}$ | ACCEPTED ESTIMATE |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | 2 | 952366 | 704951 | 726298 | XSA survivors |
| 2015 | 1 |  | 907736 | 980962 | RCT3 |
| 2016 | 0 |  |  | 980962 | GM 1957-2013 |

### 8.5 Short-term forecasts

Short-term prognoses were carried out in FLR using FLCore (2.3), projecting the stock forward three years from the 2015 (the last data year) into 2016 (the intermediate year in which the assessment is done); into 2016 (the TAC year) and finally into 2017 (the 'result' of the TAC year). For these years, a number of assumptions were made. Weight-at-age in the stock, weight-at-age in the catch and weight at age in the discards are taken to be the average over the last 3 years. The exploitation pattern (selectivity of the fishery) was taken to be the mean value of the last three years. The relative proportions of landings versus discards in the catch were taken to be the mean of the last three years.

In the intermediate year F is assumed to be equal to the estimate for F in 2015 (" F -status quo" or $\mathrm{F}_{\text {sq }}$ ). The option of assuming F to correspond to the TAC being fully landed was considered, but abandoned as an option to pursue considering the fact that the TAC has not been fully utilised in previous years. No results for this option are presented here further for that reason. Population numbers in the intermediate year for ages 2 and older are taken from the XSA survivor estimates. Numbers at age 1 in 2016 are taken from the RCT3 output and age 1 from 2017 are taken from the long-term geometric mean (1957-2013). Input to the short term forecast is presented in Table 8.5.1 and a summary of the intermediate year assumptions are given in the table below.

| ASSUMPTION | F2016 | SSB2017 | LANDINGS2016 |
| ---: | :--- | :--- | :--- |
| F2016 $=$ F2015 (Fsq) | 0.17 | 1033466 t | 109277 t |

Resulting management options for 2017 are given in Table 8.5.2.

### 8.6 Biological reference points

### 8.6.1 Precautionary approach reference points

The current precautionary approach reference points were established by the WGNSSK in 2004, when the discard estimates were included in the assessment for the first time. The stock-recruitment relationship for North Sea plaice did not show a clear breakpoint where recruitment is impaired at lower spawning stocks (Figure 8.4.2). Therefore, ICES considered that $\mathbf{B}_{\text {lim }}$ can be set at $\mathbf{B}_{\text {loss }}=160000 t$ and that $\mathbf{B}_{\text {pa }}$ can then be set at 230000 t using A multiplier of 1.44. $\mathrm{F}_{\text {lim }}$ was set at $\mathbf{F}_{\text {loss }}(0.74)$. $\mathbf{F}_{\mathbf{p a}}$ was proposed to be set at 0.6 which is the $5^{\text {th }}$ percentile of $\mathbf{F}_{\text {loss }}$ and gave a $50 \%$ probability that SSB is around $\mathbf{B}_{\mathrm{pa}}$ in the medium term. Equilibrium analysis suggests that F of 0.6 is consistent with an SSB of around 230000 t .

### 8.6.2 FMSy reference points

In 2010 ICES implemented the MSY framework for providing advice on the exploitation of stocks. The aim is to manage all stocks at an exploitation rate ( F ) that is consistent with maximum (high) long term yield while providing a low risk to the stock.
In 2014 the joint ICES-MYFISH Workshop (WKMSYREF3, ICES 2014) held place to consider the basis for FMSY ranges. The workshop was convened in response to a request from the European Commission for advice on potential intervals above and below FMSY. This resulted in an Fmsy range for North Sea plaice of $0.13-0.27$. The point value of Fmsy was set at 0.19.

This values differs from the previous value of Fmsy $=0.25$ (range $0.2-0.3$, Miller and Poos 2010).

### 8.6.3 Update of $F_{\text {lim }}$ and $F_{p a}$ values

The original $\mathrm{F}_{\mathrm{lim}}$ and $\mathrm{F}_{\mathrm{pa}}$ values were established by the WGNSSK in 2004. In 2016, an updated calculation of $\mathrm{F}_{\mathrm{lim}}$ is proposed as the F that, in equilibrium from a long-term stochastic projection, gives $50 \%$ probability of $S S B>B \lim$. The value of $F_{p a}$ is estimated as the F value such that when F is estimated to be at $\mathrm{F}_{\mathrm{pa}}$, the probability that true $\mathrm{F}<\mathrm{F}_{\text {lim }}$ is at least $95 \%$. Thus $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\lim } / \exp \left(1.645^{*} \sigma\right)$, where $\sigma$ is estimated standard deviation of $\ln (\mathrm{F})$ in the final assessment year. In case of plaice where a $\sigma$ is not available, a default value is used $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} / 1.4$. The last 10 years of the 2014 stock assessment object (data year 2004-2013) was retrieved and the distribution of recruitment at SSB was simulated using EqSIM, setting $\mathrm{B}_{\mathrm{lim}}=160000$ (Figure 8.6.1). The estimated 10 years plaice SSB are all far higher than $\mathrm{B}_{\mathrm{lim}}$. The estimated $\mathrm{F}_{\text {lim }}$ is 0.63 (Fig. 8.6.1) and the corresponding $\mathrm{F}_{\mathrm{pa}}=$ 0.45 using the default ratio of 1.4. The updated values of both $\mathrm{F}_{\lim }$ and $\mathrm{F}_{\mathrm{pa}}$ deviate from their original values, most likely due to the inclusion of Skagerrak (3.a) data in the recent years where the original reference point was not derived from. A full update of the reference points (for the stock that now includes 3.a) should be carried out in Benchmark.

### 8.7 Quality of the assessment

Although discards form a substantial part of total plaice catches, for which estimates are less certain than for landings, the assessment at present includes 14 years of discards data obtained from sampling programs in several countries and is considered to be robust and consistent between years. Discards data are now for instance available from Denmark (beamtrawls, ottertrawls, Scottish seines and Danish seines, gillnets and longliners); the United Kingdom (for beamtrawls up to 2007); Germany (beamtrawls, ottertrawls, gillnets); Belgium (beamtrawls, ottertrawls, Scottish seines) and the Netherlands (beamtrawls and ottertrawls). The improvement of retrospective patterns observed in the recent years might have beneficiated from increased coverage of discards estimates from the main fishing nations, through self-sampling and observers programs.

A self-sampling programme by the Dutch beam-trawl fleet has been in place since 2004. This sampling programme indicates spatial and temporal trends in discarding (higher discards are observed in coastal regions and late summer), but it was considered inappropriate for overall estimates of discarding because of differences in the implementations of sampling methods. In 2009, a new self-sampling programme was launched to address this. For the 2009 and 2010 assessments, discarded numbers-at-age for the Netherlands have been estimated using data from both the self-sampling and the ob-
server programmes. It is noted that estimates of discard numbers in 2010 differed between the two programmes. Mid 2011 the programme was redesigned again, to allow for better comparison between self-sampling and observer estimates through paired measurements. From 2011 onwards, Dutch discard estimates are derived exclusively from the self-sampling programme, while observer estimates are used for validation of the self-sampling data only. Preliminary analyses suggest that the self-sampling estimates are as reliable as those from the observer programme. Further analyses will be conducted in 2013 as more data from 'matched trips' (self-sampling and observer estimates from the same vessel trip) become available.

If the introduction of the landing obligation for the fisheries on sole and plaice in 2016 will affect the quality of catch data available to ICES, the quality of the assessment and advice by ICES may particularly be affected in the case of plaice, given that (substantial) discards are included in the assessment. It is unclear how these programs will continue under a landing obligation.

### 8.8 Status of the Stock

The stock is well within precautionary boundaries. SSB in 2015 is estimated around 754812 thousand tonnes which is well above Bpa ( 230000 t ). Fishing mortality in 2015 is estimated to be at a value of 0.17 (below Fpa of 0.45 , below the long term management target F of 0.30 , and below Fmsy of 0.19 ). Fishing mortality of the human consumption part of the catch is estimated to be 0.08 .

### 8.9 Management Considerations

Plaice is mainly taken by beam trawlers in a mixed fishery with sole in the southern and central part of the North Sea. There are a number of EC regulations that affect the fisheries on plaice and sole in the North Sea, e.g. as a basis for setting the TAC, limiting effort, minimum landing size and minimum mesh size.

### 8.9.1 Multiannual plan North Sea

A multiannual plan for plaice and sole in the North Sea was adopted by the EU Council in 2007 (EC regulation 676/2007). This plan is written for the North Sea stock and does not take the merging with the Skagerrak into account. The plan describes two stages: to be deemed a recovery plan during its first stage and a management plan during its second stage. ICES has evaluated this management plan in 2010 and considers it to be precautionary (ICES 2010a). Objectives are defined for these two stages; to rebuild the stocks to within safe biological limits and to exploit the stocks at MSY respectively. In 2015 WKMSYREF3 estimated Fmsy to be between 0.13 and 0.27. ICES identified the point estimate for the North Sea stock to be 0.19 (ADGMSYREF3).

Stage 1 is deemed to be completed when both stocks have been within safe biological limits for two consecutive years. The plaice stock has been within safe biological limits $(F=0.6)$ as defined by the plan since 2005. The sole stock has been within safe biological limits in terms of fishing mortality and SSB has been above the biomass limit (Bpa = 35 kt ) in the latest years. According to the management plan (Article 3.2), this signals the end of stage one. Consequently, utilisation of the plan as a basis for advice is on the basis of transitional arrangements until an evaluation of the plan has been conducted (as stipulated in article 5 of the EC regulation). In 2012, ICES evaluated a proposal by the Netherlands for an amended management plan, which could serve as the 'stage 2' plan (Coers et al. 2012). ICES concluded that the plan - subject to those amendments -
is consistent with the precautionary approach and the principle of maximum sustainable yield (MSY). However, implementation of stage two of the plan (as stipulated in article 5 of the EC regulation) is not yet defined.

Since the management plan is now in stage 2, the EU regulation stipulates that the stocks should be managed on the basis of MSY. For plaice, the ICES Fmsy estimate is 0.19 , which is below the target $\mathrm{F}(0.3)$ defined in the plan. Considering that the plan specifies that fishing mortality in stage 2 should not be below the target of 0.3 (which coincides with the upper bound of a range of Fmsy values suggested by ICES), the current advice for plaice is still on the basis of moving towards the target of 0.3 , rather than on the basis of Fmsy point estimate of 0.19 (albeit that the TAC change is restricted to a maximum $15 \%$ change). This apparent conflict in the basis for TAC setting in the management plan should be addressed.

This management plan is written for the North Sea stock. No specific management plan exists for the Skagerrak. The North Sea management plan should be updated including the Skagerrak. The forecast and advice are given for both areas with a combined TAC.

### 8.9.2 Effort regulations (North Sea)

Regulated effort restrictions in the EU were introduced in 2003 (annexes to the annual TAC regulations) for the protection of the North Sea cod stock. In addition, a long-term plan for the recovery of cod stocks was adopted in 2008 (EC regulation 1342/2008). In 2009, the effort management programme switched from a days-at-sea to a kW-day system (EC regulation 43/2009), in which different amounts of kW -days are allocated within each area by member state to different groups of vessels depending on gear and mesh size. Effort ceilings are updated annually. A minor part of the fleets exploiting sole, i.e. otter trawls (OTB) with a mesh size equal to or larger than 100 mm included in TR1, have since 2009 been affected by the regulation. The beam trawl fleet (BT2) was affected by this regulation only once in 2009 but not afterwards.

The overall fleet capacity and deployed effort of the North Sea beam trawl fleet has been substantially reduced since 1995 , likely due to a number of reasons, including the above mentioned effort limitations for the recovery of the cod stock. 25 vessels were decommissioned in 2008. In addition, the current sole and plaice long-term management plan specifically reduces effort as a management measure. However, the evaluation of amendments to the plan in 2012 showed that the plan is consistent with the precautionary approach and the principle of maximum sustainable yield (MSY) also without reductions of effort (Coers et al. 2012).
Fishing effort of the beamtrawl fleet has shifted towards the southern North Sea to target sole over the past decade. Juvenile plaice tend to be relatively abundant there, leading to relatively high discarding rates of small plaice. This shift was amongst others driven by a number of economic factors, such as the prices for sole and plaice respectively and fuel costs, which meant that the sole fishery was the most profitable fishery. With the recent substantial increases in biomass of the plaice stock, and thus to be expected increased catch rates, targeting plaice further North may become more economically favourable again. With the relatively low fishing mortality levels in recent years, it is also to be expected that a larger proportion of the population will be made up of older fish, of which the fishery could potentially benefit, since larger plaice receive higher prices on the market than small plaice. However, this benefit may be reduced if weight at age are decreasing, which seems to be the case in the plaice stock. At present, the beam trawl fleet is limited in its ability to move northwards (where
larger plaice are more abundant) by effort restrictions for the BT1 fleet, which are imposed on the basis of the North Sea cod management plan. This trade-off between objectives in the cod and flatfish plans deserves some attention. Ongoing work in the Netherlands on the levels of cod catch rates (which are considered to be low) in the beam trawl fisheries should help quantification of this trade-off. The introduction of the landing obligation will likely provide an additional strong driver for at least part of the beam trawl fleet to focus on a more northerly plaice fishery, to avoid the complications of the high unwanted bycatches of undersized plaice in the South. For effort regulations in the Skagerrak see section 07.

### 8.9.3 Technical measures

Technical measures applicable to the mixed flatfish beam-trawl fishery in the southern North Sea where sole has become relatively more abundant, affect both sole and plaice. The minimum mesh size of 80 mm selects sole at the minimum landing size. However, this mesh size generates high discards of plaice with a larger minimum landing size than sole. For the overall fleet the discards ratio has been slightly decreasing since 2003 and at present is approximately $40 \%$ by weight. Mesh enlargement would reduce the catch of undersized plaice, but would also result in loss of marketable sole. Furthermore, the size selectivity of the fleet may lead to a shift in the age and size at maturation. For example, in recent years plaice and sole have become mature at younger ages and at smaller sizes than in the past (Grift et al., 2003). The introduction of the Omega (mesh size) meter in 2010 has led to a slight increase in the effective mesh size in the fishery.

Technical management measures have caused a shift towards two categories of vessels: 2000 HP (the maximum engine power allowed) and 300 HP . The 300 HP vessels are allowed to fish within the 12-nautical mile coastal zone and in the Plaice Box. The Plaice Box is a partially closed area along the continental coast that was implemented in phases, starting in 1989. The area has been closed to most categories of vessels $>300 \mathrm{HP}$ all year round since 1995. The most recent EU-funded evaluation by Beare et al. (2010) reported the Plaice Box as having very little impact on the plaice stock.
Large scale adoption of innovative gears, for instance if EU regulations would permanently legalize the use of pulse gears could cause changes in fishing patterns in the near future (see section 8.1.3).

### 8.9.4 Frequency of assessment

The frequency of assessments was discussed at the ACOM December 2014 meeting and the Committee decided to develop simple criteria to be used to identify stocks that would be candidates for less frequent assessments. A set of four criteria were suggested based on (1) the life span of the stock, (2) stock status, (3) relative importance of recruitment in the catch forecast and (4) the quality of the assessment.
The North Sea Plaice assessment succeeded in all four criteria when evaluated in 2015 (ICES WGNSSK 2015). Therefore the North Sea Plaice stock is a candidate for less frequent assessments. The perception of the stock and the retrospective pattern in the stock did not change since last year.

### 8.10 References

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Table 8．2．1 North Sea（7．d Q1 not included）and Skagerrak Plaice．Nominal landings．

|  | North Sea |  |  |  |  |  |  |  |  |  |  |  |  | Skagerrak |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & E \\ & \stackrel{0}{6} \\ & \stackrel{\rightharpoonup}{D} \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { そ̈ } \\ & \text { డ̈ } \\ & \text { Z } \\ & 0 \end{aligned}$ |  | $\begin{aligned} & \text { む̀ } \\ & \text { む̈ } \\ & \text { む } \\ & \text { U } \end{aligned}$ | $\begin{aligned} & \text { む } \\ & \text { 芯 } \\ & \text { む } \\ & \text { む } \\ & \mathbf{Z} \end{aligned}$ | $\begin{aligned} & \text { む̀ } \\ & \text { 3 } \\ & 0 \\ & Z \end{aligned}$ | $\begin{aligned} & \text { E } \\ & 0 \\ & 0 \\ & 3 \\ & 0 \end{aligned}$ | $\frac{.}{3}$ | $\begin{gathered} \stackrel{0}{む} \\ \stackrel{1}{0} \end{gathered}$ | $\begin{aligned} & \text { त⿹\zh26灬 } \\ & \text { - } \end{aligned}$ |  | \＃ む 0 0 0 0 | $$ | $\begin{aligned} & \text { 퓽 } \\ & \hline \end{aligned}$ | $\begin{aligned} & \underset{\sim}{U} \\ & U_{1}^{\prime} \\ & \underset{H}{\prime} \end{aligned}$ |
| 1982 | 6755 | 24532 | 1046 | 3626 | 41208 | 17 | 6 | 20740 |  | 97930 | 56616 | 154546 | 140000 |  |  |
| 1983 | 9716 | 18749 | 1185 | 2397 | 51328 | 15 | 22 | 17400 |  | 100812 | 43218 | 144030 | 164000 |  |  |
| 1984 | 11393 | 22154 | 604 | 2485 | 61478 | 16 | 13 | 16853 |  | 114996 | 41153 | 156149 | 182000 | 7921 |  |
| 1985 | 9965 | 28236 | 1010 | 2197 | 90950 | 23 | 18 | 15912 |  | 148311 | 11527 | 159838 | 200000 | 10095 |  |
| 1986 | 7232 | 26332 | 751 | 1809 | 74447 | 21 | 16 | 17294 |  | 127902 | 37445 | 165347 | 180000 | 11378 |  |
| 1987 | 8554 | 21597 | 1580 | 1794 | 76612 | 12 | 7 | 20638 |  | 130794 | 22876 | 153670 | 150000 | 12503 |  |
| 1988 | 11527 | 20259 | 1773 | 2566 | 77724 | 21 | 2 | 24497 | 43 | 138412 | 16063 | 154475 | 175000 | 10820 |  |
| 1989 | 10939 | 23481 | 2037 | 5341 | 84173 | 321 | 12 | 26104 |  | 152408 | 17410 | 169818 | 185000 | 5997 |  |
| 1990 | 13940 | 26474 | 1339 | 8747 | 78204 | 1756 | 169 | 25632 |  | 156261 | －21 | 156240 | 180000 | 10048 |  |
| 1991 | 14328 | 24356 | 508 | 7926 | 67945 | 560 | 103 | 27839 |  | 143565 | 4438 | 148003 | 175000 | 6679 |  |
| 1992 | 12006 | 20891 | 537 | 6818 | 51064 | 836 | 53 | 31277 |  | 123482 | 1708 | 125190 | 175000 | 9554 | 11200 |
| 1993 | 10814 | 16452 | 603 | 6895 | 48552 | 827 | 7 | 31128 |  | 115278 | 1835 | 117113 | 175000 | 9854 | 11200 |
| 1994 | 7951 | 17056 | 407 | 5697 | 50289 | 524 | 6 | 27749 |  | 109679 | 713 | 110392 | 165000 | 9551 | 11200 |
| 1995 | 7093 | 13358 | 442 | 6329 | 44263 | 527 | 3 | 24395 |  | 96410 | 1946 | 98356 | 115000 | 9380 | 11200 |
| 1996 | 5765 | 11776 | 379 | 4780 | 35419 | 917 | 5 | 20992 |  | 80033 | 1640 | 81673 | 81000 | 8003 | 11200 |
| 1997 | 5223 | 13940 | 254 | 4159 | 34143 | 1620 | 10 | 22134 |  | 81483 | 1565 | 83048 | 91000 | 7814 | 11200 |
| 1998 | 5592 | 10087 | 489 | 2773 | 30541 | 965 | 2 | 19915 | 1 | 70365 | 1169 | 71534 | 87000 | 6449 | 11200 |
| 1999 | 6160 | 13468 | 624 | 3144 | 37513 | 643 | 4 | 17061 |  | 78617 | 2045 | 80662 | 102000 | 7049 | 11200 |
| 2000 | 7260 | 13408 | 547 | 4310 | 35030 | 883 | 3 | 20710 |  | 82151 | －1001 | 81150 | 97000 | 6989 | 11200 |
| 2001 | 6369 | 13797 | 429 | 4739 | 33290 | 1926 | 3 | 19147 |  | 79700 | 2147 | 81847 | 78000 | 9231 | 9400 |
| 2002 | 4859 | 12552 | 548 | 3927 | 29081 | 1996 | 2 | 16740 |  | 69705 | 512 | 70217 | 77000 | 7102 | 6400 |
| 2003 | 4570 | 13742 | 343 | 3800 | 27353 | 1967 | 2 | 13892 |  | 65669 | 820 | 66489 | 73250 | 7143 | 1400 |
| 2004 | 4314 | 12123 | 231 | 3649 | 23662 | 1744 | 1 | 15284 |  | 61008 | 428 | 61436 | 61000 | 8033 | 9500 |
| 2005 | 3396 | 11385 | 112 | 3379 | 22271 | 1660 | 0 | 12705 |  | 54908 | 792 | 55700 | 59000 | 6099 | 7600 |
| 2006 | 3487 | 11907 | 132 | 3599 | 22764 | 1614 | 0 | 12429 |  | 55933 | 2010 | 57943 | 57441 | 8345 | 7600 |
| 2007 | 3866 | 8128 | 144 | 2643 | 21465 | 1224 | 4 | 11557 | － | 49031 | 713 | 49744 | 50261 | 7621 | 8500 |
| 2008 | 3396 | 8229 | 125 | 3138 | 20312 | 1051 | 20 | 11411 |  | 47682 | 1193 | 48875 | 49000 | 8356 | 9300 |
| 2009 | 3474 | N／A＊ | N／A＊ | 2931 | 29142 | 1116 | 1 | 13143 | － | N／A＊ | － | 54973 | 55500 | 6514 | 9300 |
| 2010 | 3699 | 435 | 383 | 3601 | 26689 | 1089 | 5 | 14765 | － | 50666 | 10008 | 60674 | 63825 | 8700 | 9300 |
| 2011 | 4466 | 11634 | 344 | 3812 | 29272 | 1223 | 3 | 15169 | － | 65923 | 1463 | 67386 | 73400 | 8218 | 7900 |
| 2012 | 4862 | 12245 | 281 | 3742 | 32201 | 1022 | 5 | 16888 | － | 71246 | 2584 | 73830 | 84410 | 7680 | 7900 |
| 2013 | 6462 | 13650 | 249 | 4903 | 33537 | 843 | 3 | 19334 | － | 78982 | －77 | 78905 | 97070 | 6812 | 9142 |
| 2014 | 7105 | 12003 | 276 | 4203 | 29306 | 577 | 5 | 17370 | － | 69179 | 1668 | 70847 | 111631 | 9213 | 10056 |
| 2015 | 5522 | 14401 | 223 | 5171 | 32074 | 169 | 7 | 17240 | － | 74807 | 156 | 74963 | 128376 | 10480 | 10056 |
| 2016 |  |  |  |  |  |  |  |  |  |  |  |  | 131714 |  | 11766 |

＊Official estimates not available．

Table 8.2.2 Plaice in 4 and 3.a. Landings in numbers by age (including $1^{\text {st }}$ quarter of 7.d NS plaice catches) in thousands

| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1957 | 0 | 4792 | 66428 | 49659 | 35282 | 9867 | 12248 | 10026 | 5522 | 12059 |
| 1958 | 0 | 7581 | 23612 | 65979 | 36274 | 20836 | 8696 | 8507 | 6497 | 13981 |
| 1959 | 0 | 16914 | 31085 | 26040 | 41988 | 23432 | 14173 | 6547 | 6739 | 16530 |
| 1960 | 0 | 5998 | 62285 | 51359 | 21462 | 27510 | 14280 | 9073 | 5121 | 15253 |
| 1961 | 0 | 2299 | 33913 | 68965 | 33209 | 12958 | 14909 | 9900 | 6089 | 14889 |
| 1962 | 0 | 2075 | 34677 | 64548 | 48387 | 19939 | 8757 | 8733 | 5081 | 12373 |
| 1963 | 0 | 4424 | 21886 | 78412 | 55414 | 32413 | 13096 | 6965 | 7183 | 16912 |
| 1964 | 0 | 14818 | 40789 | 65219 | 57837 | 37368 | 15937 | 6644 | 4010 | 17012 |
| 1965 | 0 | 9913 | 42438 | 53486 | 43919 | 30320 | 18464 | 8602 | 4237 | 17686 |
| 1966 | 0 | 4220 | 66196 | 52428 | 37336 | 27870 | 16801 | 10981 | 6585 | 15201 |
| 1967 | 0 | 6101 | 30905 | 115157 | 42204 | 22490 | 16496 | 8163 | 6861 | 11397 |
| 1968 | 0 | 9750 | 41883 | 39251 | 127220 | 17638 | 10642 | 10396 | 4039 | 13754 |
| 1969 | 3 | 15892 | 47819 | 38185 | 37657 | 107955 | 11016 | 6440 | 8669 | 17029 |
| 1970 | 74 | 16850 | 49861 | 54712 | 39642 | 34174 | 76862 | 6149 | 4078 | 14459 |
| 1971 | 20 | 30568 | 49876 | 34580 | 26919 | 23659 | 17471 | 30711 | 6626 | 17468 |
| 1972 | 2296 | 37561 | 63958 | 54402 | 23695 | 17479 | 14787 | 11211 | 19111 | 16094 |
| 1973 | 1332 | 33342 | 62095 | 76769 | 44397 | 14517 | 9335 | 10347 | 6392 | 25194 |
| 1974 | 2305 | 23972 | 57595 | 43677 | 42588 | 20391 | 8300 | 6554 | 5773 | 22790 |
| 1975 | 1042 | 29877 | 65465 | 33211 | 27004 | 22509 | 12613 | 6292 | 4362 | 20923 |
| 1976 | 2892 | 34497 | 79621 | 98846 | 14129 | 10156 | 9352 | 6553 | 3022 | 12871 |
| 1977 | 3225 | 57061 | 43359 | 66120 | 83841 | 9157 | 5922 | 5030 | 4068 | 9206 |
| 1978 | 1102 | 58412 | 60114 | 52398 | 48310 | 34240 | 5728 | 3232 | 2333 | 7201 |
| 1979 | 1316 | 57933 | 118662 | 48879 | 47805 | 39864 | 24187 | 4154 | 2802 | 9272 |
| 1980 | 996 | 66095 | 136274 | 79035 | 25548 | 18321 | 14018 | 8621 | 1898 | 5497 |
| 1981 | 259 | 103354 | 125928 | 59565 | 36670 | 12750 | 9805 | 8295 | 5005 | 6091 |
| 1982 | 3373 | 48354 | 212188 | 71167 | 29191 | 16975 | 7704 | 5551 | 4539 | 8775 |
| 1983 | 1214 | 119696 | 115332 | 100473 | 29591 | 12960 | 8238 | 4224 | 3013 | 8308 |
| 1984 | 108 | 63507 | 280481 | 62835 | 41492 | 15417 | 6842 | 5593 | 2729 | 6551 |
| 1985 | 120 | 72806 | 146839 | 201629 | 37939 | 17106 | 7441 | 3780 | 2813 | 5830 |
| 1986 | 1669 | 66935 | 165986 | 106461 | 101684 | 27971 | 9839 | 4704 | 2834 | 7083 |
| 1987 | 1 | 85153 | 118416 | 120782 | 81304 | 44590 | 13539 | 4669 | 2346 | 5610 |
| 1988 | 1 | 15200 | 253815 | 85347 | 59950 | 31492 | 19347 | 6198 | 3434 | 6402 |
| 1989 | 1254 | 46810 | 108272 | 238243 | 58767 | 21667 | 11605 | 8025 | 2321 | 5806 |
| 1990 | 1546 | 33766 | 104796 | 119829 | 169465 | 29946 | 9053 | 4689 | 3803 | 4206 |
| 1991 | 1425 | 43064 | 87196 | 122233 | 76075 | 78728 | 15410 | 5390 | 3215 | 5634 |
| 1992 | 3386 | 43769 | 86358 | 81470 | 88534 | 37542 | 30444 | 7229 | 3295 | 6976 |
| 1993 | 3416 | 53555 | 99805 | 80856 | 63275 | 35042 | 14745 | 11500 | 3704 | 5883 |
| 1994 | 1375 | 44554 | 105863 | 86992 | 47577 | 27680 | 17279 | 6661 | 5449 | 5458 |
| 1995 | 7779 | 36761 | 82649 | 84778 | 47911 | 24572 | 14746 | 5285 | 2495 | 3896 |
| 1996 | 1103 | 43346 | 68155 | 52961 | 37285 | 19160 | 12400 | 5881 | 2799 | 4989 |
| 1997 | 897 | 43122 | 88687 | 49362 | 31750 | 18673 | 9518 | 5037 | 3054 | 4400 |
| 1998 | 197 | 30594 | 74441 | 62339 | 22793 | 9151 | 5703 | 2870 | 1983 | 3360 |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| YEAR | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ |  |  |  |  |  |
| 1999 | 549 | 8690 | 158088 | 47391 | 31778 | 14077 | 4038 | 2625 | 1597 | 3234 |  |  |  |  |  |
| 2000 | 2603 | 15656 | 40819 | 171994 | 25935 | 12586 | 2979 | 1135 | 953 | 2121 |  |  |  |  |  |
| 2001 | 4523 | 37095 | 58678 | 57195 | 101524 | 11492 | 4739 | 1212 | 650 | 2364 |  |  |  |  |  |
| 2002 | 1229 | 15868 | 60204 | 55511 | 44243 | 43066 | 6527 | 2256 | 794 | 1638 |  |  |  |  |  |
| 2003 | 700 | 44801 | 50607 | 54864 | 34689 | 20311 | 18128 | 1774 | 689 | 880 |  |  |  |  |  |
| 2004 | 544 | 12049 | 119093 | 39053 | 23766 | 13309 | 5152 | 4774 | 460 | 569 |  |  |  |  |  |
| 2005 | 2948 | 18885 | 29734 | 90989 | 20175 | 10900 | 5905 | 2760 | 2303 | 647 |  |  |  |  |  |
| 2006 | 363 | 20214 | 79934 | 34221 | 51057 | 8057 | 5589 | 2301 | 1318 | 1408 |  |  |  |  |  |
| 2007 | 1436 | 21357 | 41941 | 55949 | 20379 | 21837 | 3095 | 2011 | 604 | 1303 |  |  |  |  |  |
| 2008 | 400 | 13190 | 52382 | 45336 | 34035 | 7566 | 8066 | 978 | 735 | 936 |  |  |  |  |  |
| 2009 | 1563 | 12420 | 61907 | 42545 | 24886 | 18544 | 3400 | 4260 | 587 | 821 |  |  |  |  |  |
| 2010 | 2114 | 19874 | 49030 | 69702 | 25181 | 12622 | 9766 | 1866 | 2520 | 1267 |  |  |  |  |  |
| 2011 | 407 | 12977 | 45353 | 62017 | 51581 | 14815 | 6643 | 6984 | 1261 | 2743 |  |  |  |  |  |
| 2012 | 163 | 6164 | 60603 | 62070 | 44968 | 32037 | 7556 | 3402 | 3482 | 1924 |  |  |  |  |  |
| 2013 | 550 | 10530 | 63366 | 77056 | 42315 | 29486 | 15349 | 3955 | 2468 | 3795 |  |  |  |  |  |
| 2014 | 7 | 5384 | 40649 | 77966 | 52266 | 21932 | 12955 | 8387 | 2472 | 3440 |  |  |  |  |  |
| 2015 | 0 | 3844 | 42673 | 67065 | 60967 | 32309 | 12793 | 8902 | 4055 | 4834 |  |  |  |  |  |

Table 8.2.3 Plaice in 4 and 3.a. Discards in numbers by age (including $1^{\text {st }}$ quarter of 7.d NS plaice catches) in thousands

| AGE |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 1 | 2 | 3 | 4 | 5 | 6 |  |  |
| 1957 | 32356 | 45596 | 9220 | 909 | 961 | 25 | 0 | 0 |
| 1958 | 66199 | 73552 | 23655 | 2572 | 2137 | 65 | 0 | 0 |
| 1959 | 116086 | 127771 | 46402 | 11407 | 4737 | 106 | 0 | 0 |
| 1960 | 73939 | 167893 | 44948 | 997 | 1067 | 519 | 0 | 0 |
| 1961 | 75578 | 144609 | 89014 | 538 | 1612 | 130 | 0 | 0 |
| 1962 | 51265 | 181321 | 87599 | 21716 | 799 | 186 | 0 | 0 |
| 1963 | 90913 | 136183 | 129778 | 9964 | 2112 | 188 | 0 | 0 |
| 1964 | 66035 | 153274 | 64156 | 33825 | 3011 | 323 | 0 | 0 |
| 1965 | 43708 | 426021 | 59262 | 3404 | 923 | 267 | 0 | 0 |
| 1966 | 38496 | 163125 | 349358 | 14399 | 1402 | 125 | 0 | 0 |
| 1967 | 20199 | 133545 | 87532 | 152496 | 623 | 260 | 0 | 0 |
| 1968 | 73971 | 72192 | 46339 | 26530 | 22436 | 58 | 0 | 0 |
| 1969 | 85192 | 67378 | 16747 | 19334 | 773 | 2024 | 0 | 0 |
| 1970 | 123569 | 152480 | 27747 | 1287 | 5061 | 161 | 0 | 0 |
| 1971 | 69337 | 96968 | 42354 | 2675 | 426 | 81 | 0 | 0 |
| 1972 | 70002 | 55470 | 33899 | 5714 | 567 | 73 | 0 | 0 |
| 1973 | 132352 | 49815 | 4008 | 673 | 1289 | 67 | 0 | 0 |
| 1974 | 211139 | 308411 | 3652 | 285 | 611 | 109 | 0 | 0 |
| 1975 | 244969 | 280130 | 190536 | 4807 | 253 | 123 | 0 | 0 |
| 1976 | 183879 | 140921 | 71054 | 18013 | 174 | 41 | 0 | 0 |
| 1977 | 256628 | 103696 | 79317 | 33552 | 9317 | 129 | 0 | 0 |
| 1978 | 226872 | 154113 | 27257 | 10775 | 1244 | 570 | 0 | 0 |
| 1979 | 293166 | 215084 | 57578 | 18382 | 589 | 310 | 0 | 0 |
| 1980 | 226371 | 122561 | 932 | 687 | 193 | 86 | 0 | 0 |
| 1981 | 134142 | 193241 | 1850 | 373 | 431 | 55 | 0 | 0 |
| 1982 | 411307 | 204572 | 4624 | 1109 | 216 | 98 | 0 | 0 |
| 1983 | 261400 | 436331 | 30716 | 2235 | 804 | 72 | 0 | 0 |
| 1984 | 310675 | 313490 | 52651 | 24529 | 1492 | 69 | 0 | 0 |
| 1985 | 405385 | 229208 | 35566 | 2221 | 200 | 78 | 0 | 0 |
| 1986 | 1117345 | 490965 | 48510 | 26470 | 1451 | 146 | 0 | 0 |
| 1987 | 361519 | 1374202 | 180969 | 1427 | 1348 | 248 | 0 | 0 |
| 1988 | 348597 | 608109 | 459385 | 61167 | 882 | 177 | 0 | 0 |
| 1989 | 213291 | 485845 | 193176 | 85758 | 7224 | 115 | 0 | 0 |
| 1990 | 145314 | 279298 | 168674 | 28102 | 5011 | 177 | 0 | 0 |
| 1991 | 183126 | 301575 | 141567 | 40739 | 5528 | 939 | 0 | 0 |
| 1992 | 138755 | 219619 | 94581 | 34348 | 4307 | 880 | 0 | 0 |
| 1993 | 96371 | 154083 | 48088 | 11966 | 1635 | 216 | 0 | 0 |
| 1994 | 62122 | 95703 | 35703 | 1038 | 822 | 144 | 0 | 0 |
| 1995 | 118863 | 82676 | 15753 | 860 | 663 | 120 | 0 | 0 |
| 1996 | 111250 | 331065 | 27606 | 3930 | 451 | 116 | 0 | 0 |
| 1997 | 128653 | 510918 | 193828 | 588 | 271 | 108 | 0 | 0 |
| 1998 | 104538 | 646250 | 191631 | 53354 | 297 | 33 | 0 | 0 |


| AGE |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 1999 | 127321 | 208401 | 231769 | 54869 | 278 | 58 | 0 | 0 |
| 2000 | 103468 | 171213 | 51092 | 64971 | 1230 | 241 | 263 | 167 |
| 2001 | 30346 | 352452 | 186900 | 74744 | 54276 | 152 | 45 | 1 |
| 2002 | 310442 | 178402 | 78296 | 13940 | 2834 | 718 | 109 | 1 |
| 2003 | 67798 | 523336 | 56580 | 20184 | 4358 | 419 | 5756 | 1 |
| 2004 | 233682 | 183508 | 127876 | 10650 | 1975 | 450 | 41 | 1 |
| 2005 | 93936 | 332157 | 46454 | 23763 | 4494 | 6007 | 287 | 6 |
| 2006 | 220982 | 226944 | 117342 | 9785 | 2369 | 251 | 736 | 195 |
| 2007 | 77687 | 210407 | 73043 | 13942 | 1594 | 7028 | 190 | 1644 |
| 2008 | 135504 | 255948 | 37983 | 5356 | 1785 | 336 | 8852 | 885 |
| 2009 | 148666 | 193174 | 68975 | 9471 | 2007 | 1108 | 138 | 3220 |
| 2010 | 167387 | 180364 | 59943 | 22776 | 2699 | 1736 | 2074 | 283 |
| 2011 | 117902 | 153773 | 62696 | 37050 | 12949 | 2924 | 143 | 2273 |
| 2012 | 91961 | 313013 | 123821 | 32986 | 9439 | 1547 | 226 | 7 |
| 2013 | 128227 | 156837 | 125878 | 24797 | 4679 | 1033 | 219 | 15 |
| 2014 | 293515 | 192537 | 116178 | 55315 | 19141 | 2610 | 478 | 67 |
| 2015 | 83433 | 288990 | 130826 | 38858 | 12591 | 2367 | 521 | 209 |

Table 8.2.4 Plaice in 4 and 3.a. Catch in numbers by age (including $1^{\text {st }}$ quarter of 7.d NS plaice catches) in thousands

| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1957 | 32356 | 50388 | 75648 | 50568 | 36243 | 9892 | 12248 | 10026 | 5522 | 12059 |
| 1958 | 66199 | 81133 | 47267 | 68551 | 38411 | 20901 | 8696 | 8507 | 6497 | 13981 |
| 1959 | 116086 | 144685 | 77487 | 37447 | 46725 | 23538 | 14173 | 6547 | 6739 | 16530 |
| 1960 | 73939 | 173891 | 107233 | 52356 | 22529 | 28029 | 14280 | 9073 | 5121 | 15253 |
| 1961 | 75578 | 146908 | 122927 | 69503 | 34821 | 13088 | 14909 | 9900 | 6089 | 14889 |
| 1962 | 51265 | 183396 | 122276 | 86264 | 49186 | 20125 | 8757 | 8733 | 5081 | 12373 |
| 1963 | 90913 | 140607 | 151664 | 88376 | 57526 | 32601 | 13096 | 6965 | 7183 | 16912 |
| 1964 | 66035 | 168092 | 104945 | 99044 | 60848 | 37691 | 15937 | 6644 | 4010 | 17012 |
| 1965 | 43708 | 435934 | 101700 | 56890 | 44842 | 30587 | 18464 | 8602 | 4237 | 17686 |
| 1966 | 38496 | 167345 | 415554 | 66827 | 38738 | 27995 | 16801 | 10981 | 6585 | 15201 |
| 1967 | 20199 | 139646 | 118437 | 267653 | 42827 | 22750 | 16496 | 8163 | 6861 | 11397 |
| 1968 | 73971 | 81942 | 88222 | 65781 | 149656 | 17696 | 10642 | 10396 | 4039 | 13754 |
| 1969 | 85195 | 83270 | 64566 | 57519 | 38430 | 109979 | 11016 | 6440 | 8669 | 17029 |
| 1970 | 123643 | 169330 | 77608 | 55999 | 44703 | 34335 | 76862 | 6149 | 4078 | 14459 |
| 1971 | 69357 | 127536 | 92230 | 37255 | 27345 | 23740 | 17471 | 30711 | 6626 | 17468 |
| 1972 | 72298 | 93031 | 97857 | 60116 | 24262 | 17552 | 14787 | 11211 | 19111 | 16094 |
| 1973 | 133684 | 83157 | 66103 | 77442 | 45686 | 14584 | 9335 | 10347 | 6392 | 25194 |
| 1974 | 213444 | 332383 | 61247 | 43962 | 43199 | 20500 | 8300 | 6554 | 5773 | 22790 |
| 1975 | 246011 | 310007 | 256001 | 38018 | 27257 | 22632 | 12613 | 6292 | 4362 | 20923 |
| 1976 | 186771 | 175418 | 150675 | 116859 | 14303 | 10197 | 9352 | 6553 | 3022 | 12871 |
| 1977 | 259853 | 160757 | 122676 | 99672 | 93158 | 9286 | 5922 | 5030 | 4068 | 9206 |
| 1978 | 227974 | 212525 | 87371 | 63173 | 49554 | 34810 | 5728 | 3232 | 2333 | 7201 |
| 1979 | 294482 | 273017 | 176240 | 67261 | 48394 | 40174 | 24187 | 4154 | 2802 | 9272 |
| 1980 | 227367 | 188656 | 137206 | 79722 | 25741 | 18407 | 14018 | 8621 | 1898 | 5497 |
| 1981 | 134401 | 296595 | 127778 | 59938 | 37101 | 12805 | 9805 | 8295 | 5005 | 6091 |
| 1982 | 414680 | 252926 | 216812 | 72276 | 29407 | 17073 | 7704 | 5551 | 4539 | 8775 |
| 1983 | 262614 | 556027 | 146048 | 102708 | 30395 | 13032 | 8238 | 4224 | 3013 | 8308 |
| 1984 | 310783 | 376997 | 333132 | 87364 | 42984 | 15486 | 6842 | 5593 | 2729 | 6551 |
| 1985 | 405505 | 302014 | 182405 | 203850 | 38139 | 17184 | 7441 | 3780 | 2813 | 5830 |
| 1986 | 1119014 | 557900 | 214496 | 132931 | 103135 | 28117 | 9839 | 4704 | 2834 | 7083 |
| 1987 | 361520 | 1459355 | 299385 | 122209 | 82652 | 44838 | 13539 | 4669 | 2346 | 5610 |
| 1988 | 348598 | 623309 | 713200 | 146514 | 60832 | 31669 | 19347 | 6198 | 3434 | 6402 |
| 1989 | 214545 | 532655 | 301448 | 324001 | 65991 | 21782 | 11605 | 8025 | 2321 | 5806 |
| 1990 | 146860 | 313064 | 273470 | 147931 | 174476 | 30123 | 9053 | 4689 | 3803 | 4206 |
| 1991 | 184551 | 344639 | 228763 | 162972 | 81603 | 79667 | 15410 | 5390 | 3215 | 5634 |
| 1992 | 142141 | 263388 | 180939 | 115818 | 92841 | 38422 | 30444 | 7229 | 3295 | 6976 |
| 1993 | 99787 | 207638 | 147893 | 92822 | 64910 | 35258 | 14745 | 11500 | 3704 | 5883 |
| 1994 | 63497 | 140257 | 141566 | 88030 | 48399 | 27824 | 17279 | 6661 | 5449 | 5458 |
| 1995 | 126642 | 119437 | 98402 | 85638 | 48574 | 24692 | 14746 | 5285 | 2495 | 3896 |
| 1996 | 112353 | 374411 | 95761 | 56891 | 37736 | 19276 | 12400 | 5881 | 2799 | 4989 |
| 1997 | 129550 | 554040 | 282515 | 49950 | 32021 | 18781 | 9518 | 5037 | 3054 | 4400 |
| 1998 | 104735 | 676844 | 266072 | 115693 | 23090 | 9184 | 5703 | 2870 | 1983 | 3360 |


| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1999 | 127870 | 217091 | 389857 | 102260 | 32056 | 14135 | 4038 | 2625 | 1597 | 3234 |
| 2000 | 106071 | 186869 | 91911 | 236965 | 27165 | 12827 | 3242 | 1302 | 953 | 2121 |
| 2001 | 34869 | 389547 | 245578 | 131939 | 155800 | 11644 | 4784 | 1213 | 650 | 2364 |
| 2002 | 311671 | 194270 | 138500 | 69451 | 47077 | 43784 | 6636 | 2257 | 794 | 1638 |
| 2003 | 68498 | 568137 | 107187 | 75048 | 39047 | 20730 | 23884 | 1775 | 689 | 880 |
| 2004 | 234226 | 195557 | 246969 | 49703 | 25741 | 13759 | 5193 | 4775 | 460 | 569 |
| 2005 | 96884 | 351042 | 76188 | 114752 | 24669 | 16907 | 6192 | 2766 | 2303 | 647 |
| 2006 | 221345 | 247158 | 197276 | 44006 | 53426 | 8308 | 6325 | 2496 | 1318 | 1408 |
| 2007 | 79123 | 231764 | 114984 | 69891 | 21973 | 28865 | 3285 | 3655 | 604 | 1303 |
| 2008 | 135904 | 269138 | 90365 | 50692 | 35820 | 7902 | 16918 | 1863 | 735 | 936 |
| 2009 | 150229 | 205594 | 130882 | 52016 | 26893 | 19652 | 3538 | 7480 | 587 | 821 |
| 2010 | 169501 | 200238 | 108973 | 92478 | 27880 | 14358 | 11840 | 2149 | 2520 | 1267 |
| 2011 | 118309 | 166750 | 108049 | 99067 | 64530 | 17739 | 6786 | 9257 | 1261 | 2743 |
| 2012 | 92124 | 319177 | 184424 | 95056 | 54407 | 33584 | 7782 | 3409 | 3482 | 1924 |
| 2013 | 128777 | 167367 | 189244 | 101853 | 46994 | 30519 | 15568 | 3970 | 2468 | 3795 |
| 2014 | 293522 | 197921 | 156827 | 133281 | 71407 | 24542 | 13433 | 8454 | 2472 | 3440 |
| 2015 | 83433 | 292834 | 173499 | 105923 | 73558 | 34676 | 13314 | 9111 | 4055 | 4834 |

Table 8.2.5 Plaice in 4 and 3.a. Stock weight at age (kg).

| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1957 | 0.038 | 0.102 | 0.157 | 0.242 | 0.325 | 0.485 | 0.719 | 0.682 | 0.844 | 0.918 |
| 1958 | 0.041 | 0.093 | 0.180 | 0.272 | 0.303 | 0.442 | 0.577 | 0.778 | 0.793 | 0.945 |
| 1959 | 0.045 | 0.106 | 0.173 | 0.264 | 0.329 | 0.470 | 0.650 | 0.686 | 0.908 | 0.897 |
| 1960 | 0.038 | 0.111 | 0.181 | 0.272 | 0.364 | 0.469 | 0.633 | 0.726 | 0.845 | 0.918 |
| 1961 | 0.037 | 0.098 | 0.185 | 0.306 | 0.337 | 0.483 | 0.579 | 0.691 | 0.779 | 0.911 |
| 1962 | 0.036 | 0.096 | 0.173 | 0.301 | 0.424 | 0.573 | 0.684 | 0.806 | 0.873 | 1.335 |
| 1963 | 0.041 | 0.103 | 0.176 | 0.273 | 0.378 | 0.540 | 0.663 | 0.788 | 0.882 | 0.961 |
| 1964 | 0.024 | 0.113 | 0.184 | 0.296 | 0.373 | 0.477 | 0.645 | 0.673 | 0.845 | 0.973 |
| 1965 | 0.031 | 0.068 | 0.198 | 0.294 | 0.333 | 0.43 | 0.516 | 0.601 | 0.722 | 0.578 |
| 1966 | 0.031 | 0.099 | 0.127 | 0.305 | 0.403 | 0.455 | 0.503 | 0.565 | 0.581 | 0.848 |
| 1967 | 0.029 | 0.104 | 0.179 | 0.205 | 0.442 | 0.528 | 0.585 | 0.650 | 0.703 | 0.833 |
| 1968 | 0.055 | 0.094 | 0.175 | 0.287 | 0.344 | 0.532 | 0.592 | 0.362 | 0.667 | 0.746 |
| 1969 | 0.047 | 0.158 | 0.188 | 0.266 | 0.344 | 0.390 | 0.565 | 0.621 | 0.679 | 0.635 |
| 1970 | 0.043 | 0.113 | 0.236 | 0.274 | 0.369 | 0.410 | 0.468 | 0.636 | 0.732 | 0.747 |
| 1971 | 0.051 | 0.109 | 0.251 | 0.344 | 0.413 | 0.489 | 0.512 | 0.583 | 0.696 | 0.707 |
| 1972 | 0.056 | 0.158 | 0.218 | 0.407 | 0.473 | 0.534 | 0.579 | 0.606 | 0.655 | 0.759 |
| 1973 | 0.037 | 0.134 | 0.237 | 0.308 | 0.468 | 0.521 | 0.566 | 0.583 | 0.617 | 0.690 |
| 1974 | 0.049 | 0.105 | 0.217 | 0.416 | 0.437 | 0.524 | 0.570 | 0.629 | 0.652 | 0.690 |
| 1975 | 0.063 | 0.141 | 0.187 | 0.388 | 0.483 | 0.544 | 0.610 | 0.668 | 0.704 | 0.762 |
| 1976 | 0.082 | 0.169 | 0.226 | 0.308 | 0.484 | 0.550 | 0.593 | 0.658 | 0.694 | 0.743 |
| 1977 | 0.064 | 0.184 | 0.265 | 0.311 | 0.405 | 0.551 | 0.627 | 0.690 | 0.667 | 0.759 |
| 1978 | 0.064 | 0.151 | 0.319 | 0.373 | 0.411 | 0.467 | 0.547 | 0.630 | 0.704 | 0.773 |
| 1979 | 0.062 | 0.179 | 0.258 | 0.365 | 0.414 | 0.459 | 0.543 | 0.667 | 0.764 | 0.826 |
| 1980 | 0.049 | 0.163 | 0.289 | 0.428 | 0.444 | 0.524 | 0.582 | 0.651 | 0.778 | 1.025 |
| 1981 | 0.041 | 0.140 | 0.239 | 0.421 | 0.473 | 0.536 | 0.570 | 0.624 | 0.707 | 0.849 |
| 1982 | 0.048 | 0.128 | 0.250 | 0.351 | 0.490 | 0.589 | 0.631 | 0.679 | 0.726 | 0.828 |
| 1983 | 0.045 | 0.128 | 0.242 | 0.381 | 0.494 | 0.559 | 0.624 | 0.712 | 0.754 | 0.791 |
| 1984 | 0.048 | 0.129 | 0.216 | 0.413 | 0.464 | 0.571 | 0.649 | 0.692 | 0.787 | 0.898 |
| 1985 | 0.048 | 0.146 | 0.232 | 0.320 | 0.452 | 0.536 | 0.635 | 0.656 | 0.764 | 0.869 |
| 1986 | 0.043 | 0.126 | 0.245 | 0.311 | 0.440 | 0.533 | 0.692 | 0.779 | 0.888 | 0.971 |
| 1987 | 0.036 | 0.105 | 0.200 | 0.383 | 0.401 | 0.503 | 0.573 | 0.711 | 0.747 | 0.817 |
| 1988 | 0.036 | 0.097 | 0.172 | 0.264 | 0.426 | 0.467 | 0.547 | 0.644 | 0.706 | 0.897 |
| 1989 | 0.039 | 0.101 | 0.192 | 0.247 | 0.362 | 0.484 | 0.553 | 0.616 | 0.759 | 0.837 |
| 1990 | 0.043 | 0.108 | 0.176 | 0.261 | 0.343 | 0.422 | 0.555 | 0.647 | 0.701 | 0.760 |
| 1991 | 0.048 | 0.131 | 0.184 | 0.260 | 0.342 | 0.401 | 0.463 | 0.633 | 0.652 | 0.744 |
| 1992 | 0.043 | 0.121 | 0.199 | 0.270 | 0.318 | 0.403 | 0.500 | 0.573 | 0.683 | 0.730 |
| 1993 | 0.050 | 0.119 | 0.208 | 0.315 | 0.330 | 0.391 | 0.490 | 0.587 | 0.633 | 0.723 |
| 1994 | 0.053 | 0.141 | 0.214 | 0.290 | 0.360 | 0.404 | 0.462 | 0.533 | 0.653 | 0.702 |
| 1995 | 0.050 | 0.142 | 0.254 | 0.336 | 0.399 | 0.448 | 0.509 | 0.584 | 0.678 | 0.789 |
| 1996 | 0.044 | 0.117 | 0.229 | 0.368 | 0.390 | 0.462 | 0.488 | 0.554 | 0.660 | 0.791 |
| 1997 | 0.035 | 0.115 | 0.233 | 0.359 | 0.439 | 0.492 | 0.521 | 0.543 | 0.627 | 0.734 |
| 1998 | 0.038 | 0.081 | 0.207 | 0.333 | 0.474 | 0.577 | 0.581 | 0.648 | 0.656 | 0.642 |
| 1999 | 0.044 | 0.091 | 0.150 | 0.319 | 0.437 | 0.524 | 0.586 | 0.644 | 0.664 | 0.620 |


| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ |
| 2000 | 0.051 | 0.106 | 0.165 | 0.219 | 0.408 | 0.467 | 0.649 | 0.695 | 0.656 | 0.744 |
| 2001 | 0.061 | 0.122 | 0.202 | 0.233 | 0.331 | 0.452 | 0.560 | 0.641 | 0.798 | 0.816 |
| 2002 | 0.048 | 0.118 | 0.213 | 0.301 | 0.319 | 0.403 | 0.446 | 0.612 | 0.685 | 0.781 |
| 2003 | 0.057 | 0.111 | 0.227 | 0.269 | 0.344 | 0.391 | 0.464 | 0.600 | 0.714 | 0.960 |
| 2004 | 0.047 | 0.116 | 0.201 | 0.306 | 0.384 | 0.430 | 0.489 | 0.495 | 0.780 | 0.921 |
| 2005 | 0.053 | 0.106 | 0.216 | 0.237 | 0.378 | 0.422 | 0.434 | 0.527 | 0.621 | 0.815 |
| 2006 | 0.052 | 0.130 | 0.190 | 0.316 | 0.354 | 0.424 | 0.439 | 0.506 | 0.583 | 0.688 |
| 2007 | 0.047 | 0.093 | 0.235 | 0.238 | 0.337 | 0.394 | 0.458 | 0.412 | 0.526 | 0.512 |
| 2008 | 0.048 | 0.114 | 0.196 | 0.274 | 0.355 | 0.429 | 0.484 | 0.627 | 0.598 | 0.449 |
| 2009 | 0.052 | 0.114 | 0.194 | 0.344 | 0.373 | 0.412 | 0.472 | 0.540 | 0.565 | 0.576 |
| 2010 | 0.053 | 0.116 | 0.179 | 0.340 | 0.361 | 0.401 | 0.448 | 0.572 | 0.568 | 0.655 |
| 2011 | 0.039 | 0.100 | 0.187 | 0.209 | 0.355 | 0.483 | 0.438 | 0.422 | 0.530 | 0.580 |
| 2012 | 0.052 | 0.093 | 0.142 | 0.188 | 0.331 | 0.393 | 0.484 | 0.479 | 0.480 | 0.518 |
| 2013 | 0.043 | 0.107 | 0.153 | 0.208 | 0.320 | 0.354 | 0.434 | 0.493 | 0.662 | 0.468 |
| 2014 | 0.048 | 0.104 | 0.158 | 0.202 | 0.312 | 0.380 | 0.439 | 0.484 | 0.458 | 0.615 |
| 2015 | 0.024 | 0.065 | 0.120 | 0.207 | 0.279 | 0.323 | 0.379 | 0.435 | 0.465 | 0.457 |

Table 8.2.6 Plaice in 4 and 3.a. Landings weight at age (kg).

| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1957 | 0.000 | 0.165 | 0.201 | 0.258 | 0.353 | 0.456 | 0.533 | 0.589 | 0.396 | 0.998 |
| 1958 | 0.000 | 0.198 | 0.221 | 0.259 | 0.337 | 0.453 | 0.513 | 0.615 | 0.665 | 0.992 |
| 1959 | 0.000 | 0.218 | 0.246 | 0.293 | 0.362 | 0.473 | 0.592 | 0.623 | 0.750 | 1.000 |
| 1960 | 0.000 | 0.200 | 0.236 | 0.289 | 0.386 | 0.485 | 0.601 | 0.683 | 0.724 | 1.094 |
| 1961 | 0.000 | 0.191 | 0.233 | 0.302 | 0.412 | 0.509 | 0.604 | 0.671 | 0.812 | 1.071 |
| 1962 | 0.000 | 0.211 | 0.248 | 0.300 | 0.400 | 0.541 | 0.570 | 0.692 | 0.777 | 1.127 |
| 1963 | 0.000 | 0.253 | 0.286 | 0.319 | 0.399 | 0.533 | 0.624 | 0.667 | 0.715 | 1.028 |
| 1964 | 0.000 | 0.250 | 0.273 | 0.312 | 0.388 | 0.487 | 0.628 | 0.700 | 0.737 | 1.005 |
| 1965 | 0.000 | 0.242 | 0.282 | 0.321 | 0.385 | 0.471 | 0.539 | 0.663 | 0.726 | 0.887 |
| 1966 | 0.000 | 0.232 | 0.270 | 0.348 | 0.436 | 0.484 | 0.559 | 0.624 | 0.690 | 0.933 |
| 1967 | 0.000 | 0.232 | 0.279 | 0.322 | 0.425 | 0.547 | 0.597 | 0.662 | 0.738 | 0.978 |
| 1968 | 0.000 | 0.267 | 0.298 | 0.331 | 0.366 | 0.517 | 0.590 | 0.596 | 0.686 | 0.911 |
| 1969 | 0.217 | 0.294 | 0.310 | 0.333 | 0.359 | 0.412 | 0.573 | 0.655 | 0.658 | 0.893 |
| 1970 | 0.315 | 0.286 | 0.318 | 0.356 | 0.419 | 0.443 | 0.499 | 0.672 | 0.744 | 0.892 |
| 1971 | 0.256 | 0.318 | 0.356 | 0.403 | 0.448 | 0.514 | 0.542 | 0.607 | 0.699 | 0.891 |
| 1972 | 0.246 | 0.296 | 0.352 | 0.428 | 0.493 | 0.541 | 0.608 | 0.646 | 0.674 | 0.939 |
| 1973 | 0.272 | 0.316 | 0.344 | 0.405 | 0.486 | 0.539 | 0.605 | 0.627 | 0.677 | 0.842 |
| 1974 | 0.285 | 0.311 | 0.354 | 0.405 | 0.476 | 0.554 | 0.609 | 0.693 | 0.707 | 0.926 |
| 1975 | 0.249 | 0.300 | 0.330 | 0.420 | 0.495 | 0.587 | 0.636 | 0.703 | 0.783 | 1.019 |
| 1976 | 0.265 | 0.295 | 0.338 | 0.375 | 0.513 | 0.594 | 0.641 | 0.705 | 0.741 | 0.980 |
| 1977 | 0.254 | 0.323 | 0.353 | 0.380 | 0.418 | 0.556 | 0.647 | 0.721 | 0.715 | 0.978 |
| 1978 | 0.244 | 0.315 | 0.369 | 0.397 | 0.438 | 0.491 | 0.609 | 0.687 | 0.776 | 0.950 |
| 1979 | 0.235 | 0.311 | 0.349 | 0.388 | 0.429 | 0.474 | 0.550 | 0.675 | 0.796 | 0.960 |
| 1980 | 0.238 | 0.286 | 0.344 | 0.401 | 0.473 | 0.545 | 0.588 | 0.662 | 0.772 | 1.013 |
| 1981 | 0.237 | 0.274 | 0.329 | 0.416 | 0.505 | 0.558 | 0.604 | 0.642 | 0.725 | 1.007 |
| 1982 | 0.279 | 0.262 | 0.311 | 0.424 | 0.514 | 0.608 | 0.664 | 0.712 | 0.738 | 0.984 |
| 1983 | 0.200 | 0.250 | 0.300 | 0.383 | 0.515 | 0.604 | 0.677 | 0.771 | 0.815 | 0.984 |
| 1984 | 0.231 | 0.263 | 0.283 | 0.364 | 0.480 | 0.591 | 0.677 | 0.726 | 0.839 | 1.036 |
| 1985 | 0.245 | 0.264 | 0.290 | 0.335 | 0.445 | 0.563 | 0.667 | 0.730 | 0.807 | 1.021 |
| 1986 | 0.221 | 0.269 | 0.303 | 0.339 | 0.405 | 0.473 | 0.668 | 0.750 | 0.856 | 1.014 |
| 1987 | 0.000 | 0.249 | 0.299 | 0.345 | 0.378 | 0.472 | 0.574 | 0.728 | 0.835 | 0.993 |
| 1988 | 0.000 | 0.254 | 0.278 | 0.341 | 0.418 | 0.478 | 0.590 | 0.680 | 0.808 | 1.017 |
| 1989 | 0.236 | 0.280 | 0.308 | 0.331 | 0.385 | 0.515 | 0.591 | 0.668 | 0.785 | 0.940 |
| 1990 | 0.271 | 0.284 | 0.297 | 0.315 | 0.364 | 0.441 | 0.586 | 0.690 | 0.761 | 1.010 |
| 1991 | 0.227 | 0.286 | 0.292 | 0.302 | 0.360 | 0.452 | 0.526 | 0.666 | 0.743 | 0.924 |
| 1992 | 0.251 | 0.263 | 0.290 | 0.312 | 0.330 | 0.415 | 0.530 | 0.607 | 0.719 | 0.891 |
| 1993 | 0.249 | 0.273 | 0.288 | 0.319 | 0.343 | 0.408 | 0.512 | 0.630 | 0.720 | 0.856 |
| 1994 | 0.229 | 0.263 | 0.284 | 0.333 | 0.375 | 0.417 | 0.491 | 0.610 | 0.731 | 0.906 |
| 1995 | 0.272 | 0.277 | 0.301 | 0.335 | 0.375 | 0.420 | 0.474 | 0.593 | 0.734 | 0.906 |
| 1996 | 0.240 | 0.279 | 0.304 | 0.346 | 0.415 | 0.465 | 0.490 | 0.553 | 0.712 | 0.858 |
| 1997 | 0.208 | 0.271 | 0.313 | 0.355 | 0.410 | 0.474 | 0.541 | 0.574 | 0.616 | 0.912 |
| 1998 | 0.151 | 0.260 | 0.306 | 0.384 | 0.452 | 0.546 | 0.613 | 0.673 | 0.687 | 0.899 |
| 1999 | 0.245 | 0.253 | 0.280 | 0.347 | 0.415 | 0.416 | 0.538 | 0.637 | 0.748 | 0.804 |


| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ |
| 2000 | 0.228 | 0.267 | 0.283 | 0.312 | 0.378 | 0.461 | 0.597 | 0.689 | 0.752 | 0.888 |
| 2001 | 0.238 | 0.267 | 0.291 | 0.307 | 0.360 | 0.412 | 0.582 | 0.701 | 0.796 | 0.799 |
| 2002 | 0.237 | 0.264 | 0.289 | 0.311 | 0.336 | 0.430 | 0.477 | 0.644 | 0.760 | 0.904 |
| 2003 | 0.232 | 0.252 | 0.285 | 0.320 | 0.353 | 0.389 | 0.482 | 0.635 | 0.763 | 0.857 |
| 2004 | 0.214 | 0.246 | 0.281 | 0.328 | 0.391 | 0.429 | 0.508 | 0.560 | 0.797 | 0.872 |
| 2005 | 0.272 | 0.265 | 0.280 | 0.330 | 0.382 | 0.426 | 0.465 | 0.555 | 0.617 | 0.910 |
| 2006 | 0.253 | 0.267 | 0.282 | 0.322 | 0.383 | 0.389 | 0.457 | 0.477 | 0.531 | 0.748 |
| 2007 | 0.263 | 0.268 | 0.303 | 0.343 | 0.364 | 0.432 | 0.507 | 0.486 | 0.587 | 0.632 |
| 2008 | 0.249 | 0.269 | 0.309 | 0.341 | 0.400 | 0.446 | 0.531 | 0.720 | 0.640 | 0.638 |
| 2009 | 0.176 | 0.260 | 0.308 | 0.355 | 0.415 | 0.481 | 0.531 | 0.608 | 0.668 | 0.792 |
| 2010 | 0.206 | 0.265 | 0.308 | 0.348 | 0.418 | 0.476 | 0.516 | 0.625 | 0.682 | 0.649 |
| 2011 | 0.235 | 0.242 | 0.281 | 0.341 | 0.414 | 0.504 | 0.604 | 0.521 | 0.556 | 0.804 |
| 2012 | 0.236 | 0.258 | 0.305 | 0.351 | 0.380 | 0.436 | 0.518 | 0.558 | 0.558 | 0.680 |
| 2013 | 0.031 | 0.242 | 0.281 | 0.313 | 0.364 | 0.417 | 0.494 | 0.600 | 0.607 | 0.680 |
| 2014 | 0.207 | 0.252 | 0.285 | 0.318 | 0.368 | 0.418 | 0.479 | 0.543 | 0.628 | 0.650 |
| 2015 | NA | 0.251 | 0.284 | 0.321 | 0.359 | 0.409 | 0.473 | 0.487 | 0.582 | 0.600 |

Table 8.2.7 Plaice in 4 and 3.a. Discards weight at age (kg).

| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1957 | 0.044 | 0.104 | 0.146 | 0.181 | 0.206 | 0.244 | 0.244 | 0.231 | 0.000 | 0.000 |
| 1958 | 0.047 | 0.096 | 0.158 | 0.188 | 0.200 | 0.244 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1959 | 0.051 | 0.107 | 0.155 | 0.186 | 0.197 | 0.231 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1960 | 0.045 | 0.112 | 0.159 | 0.188 | 0.204 | 0.212 | 0.244 | 0.000 | 0.000 | 0.000 |
| 1961 | 0.044 | 0.100 | 0.160 | 0.194 | 0.204 | 0.220 | 0.220 | 0.000 | 0.000 | 0.000 |
| 1962 | 0.042 | 0.098 | 0.155 | 0.193 | 0.213 | 0.221 | 0.221 | 0.231 | 0.000 | 0.000 |
| 1963 | 0.048 | 0.105 | 0.156 | 0.188 | 0.205 | 0.231 | 0.221 | 0.231 | 0.000 | 0.000 |
| 1964 | 0.032 | 0.114 | 0.160 | 0.192 | 0.204 | 0.221 | 0.244 | 0.231 | 0.000 | 0.000 |
| 1965 | 0.038 | 0.072 | 0.166 | 0.192 | 0.212 | 0.221 | 0.231 | 0.000 | 0.000 | 0.000 |
| 1966 | 0.038 | 0.101 | 0.125 | 0.194 | 0.205 | 0.231 | 0.231 | 0.244 | 0.000 | 0.000 |
| 1967 | 0.036 | 0.105 | 0.158 | 0.169 | 0.220 | 0.220 | 0.244 | 0.244 | 0.000 | 0.000 |
| 1968 | 0.060 | 0.096 | 0.156 | 0.191 | 0.192 | 0.244 | 0.220 | 0.000 | 0.000 | 0.000 |
| 1969 | 0.052 | 0.146 | 0.162 | 0.186 | 0.211 | 0.212 | 0.000 | 0.231 | 0.000 | 0.000 |
| 1970 | 0.049 | 0.114 | 0.179 | 0.189 | 0.196 | 0.000 | 0.220 | 0.231 | 0.000 | 0.000 |
| 1971 | 0.057 | 0.110 | 0.183 | 0.200 | 0.212 | 0.000 | 0.000 | 0.231 | 0.000 | 0.000 |
| 1972 | 0.061 | 0.147 | 0.173 | 0.211 | 0.211 | 0.244 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1973 | 0.043 | 0.131 | 0.179 | 0.195 | 0.211 | 0.244 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1974 | 0.054 | 0.106 | 0.173 | 0.212 | 0.220 | 0.231 | 0.244 | 0.000 | 0.000 | 0.000 |
| 1975 | 0.068 | 0.136 | 0.162 | 0.206 | 0.221 | 0.244 | 0.244 | 0.000 | 0.000 | 0.000 |
| 1976 | 0.085 | 0.153 | 0.176 | 0.195 | 0.220 | 0.000 | 0.244 | 0.000 | 0.000 | 0.000 |
| 1977 | 0.069 | 0.160 | 0.186 | 0.196 | 0.198 | 0.220 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1978 | 0.069 | 0.143 | 0.197 | 0.205 | 0.211 | 0.213 | 0.231 | 0.000 | 0.000 | 0.000 |
| 1979 | 0.066 | 0.158 | 0.185 | 0.204 | 0.220 | 0.231 | 0.221 | 0.244 | 0.000 | 0.000 |
| 1980 | 0.055 | 0.149 | 0.191 | 0.212 | 0.231 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1981 | 0.048 | 0.135 | 0.179 | 0.212 | 0.220 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1982 | 0.054 | 0.126 | 0.182 | 0.203 | 0.231 | 0.244 | 0.244 | 0.000 | 0.000 | 0.000 |
| 1983 | 0.051 | 0.126 | 0.180 | 0.205 | 0.211 | 0.244 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1984 | 0.053 | 0.127 | 0.172 | 0.211 | 0.205 | 0.000 | 0.244 | 0.000 | 0.000 | 0.000 |
| 1985 | 0.054 | 0.139 | 0.177 | 0.197 | 0.231 | 0.244 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1986 | 0.049 | 0.124 | 0.181 | 0.196 | 0.220 | 0.244 | 0.244 | 0.000 | 0.000 | 0.000 |
| 1987 | 0.043 | 0.105 | 0.166 | 0.205 | 0.220 | 0.231 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1988 | 0.043 | 0.098 | 0.153 | 0.185 | 0.220 | 0.244 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1989 | 0.046 | 0.102 | 0.163 | 0.181 | 0.196 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.051 | 0.111 | 0.157 | 0.186 | 0.212 | 0.231 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1991 | 0.055 | 0.130 | 0.161 | 0.185 | 0.203 | 0.221 | 0.231 | 0.231 | 0.000 | 0.000 |
| 1992 | 0.050 | 0.122 | 0.167 | 0.188 | 0.204 | 0.212 | 0.231 | 0.244 | 0.000 | 0.000 |
| 1993 | 0.056 | 0.121 | 0.171 | 0.197 | 0.211 | 0.231 | 0.244 | 0.000 | 0.000 | 0.000 |
| 1994 | 0.060 | 0.140 | 0.175 | 0.194 | 0.213 | 0.244 | 0.244 | 0.221 | 0.000 | 0.000 |
| 1995 | 0.058 | 0.141 | 0.186 | 0.201 | 0.220 | 0.232 | 0.232 | 0.244 | 0.000 | 0.000 |
| 1996 | 0.052 | 0.122 | 0.179 | 0.205 | 0.221 | 0.232 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.044 | 0.117 | 0.178 | 0.203 | 0.221 | 0.244 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1998 | 0.047 | 0.086 | 0.170 | 0.199 | 0.220 | 0.000 | 0.244 | 0.000 | 0.000 | 0.000 |
| 1999 | 0.053 | 0.097 | 0.143 | 0.197 | 0.220 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |


| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ |
| 2000 | 0.059 | 0.110 | 0.151 | 0.174 | 0.244 | 0.000 | 0.203 | 0.000 | 0.000 | 0.000 |
| 2001 | 0.068 | 0.122 | 0.167 | 0.178 | 0.197 | 0.244 | 0.000 | 0.244 | 0.000 | 0.000 |
| 2002 | 0.056 | 0.119 | 0.170 | 0.182 | 0.172 | 0.208 | 0.003 | 0.000 | 0.000 | 0.000 |
| 2003 | 0.064 | 0.113 | 0.174 | 0.185 | 0.198 | 0.204 | 0.221 | 0.000 | 0.000 | 0.000 |
| 2004 | 0.054 | 0.117 | 0.164 | 0.183 | 0.189 | 0.192 | 0.196 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.061 | 0.109 | 0.170 | 0.175 | 0.215 | 0.205 | 0.210 | 0.176 | 0.000 | 0.000 |
| 2006 | 0.060 | 0.128 | 0.164 | 0.193 | 0.198 | 0.204 | 0.212 | 0.220 | 0.000 | 0.000 |
| 2007 | 0.055 | 0.098 | 0.177 | 0.178 | 0.188 | 0.199 | 0.225 | 0.200 | 0.000 | 0.000 |
| 2008 | 0.056 | 0.116 | 0.163 | 0.186 | 0.187 | 0.230 | 0.220 | 0.191 | 0.000 | 0.000 |
| 2009 | 0.060 | 0.116 | 0.164 | 0.199 | 0.202 | 0.212 | 0.210 | 0.220 | 0.000 | 0.000 |
| 2010 | 0.060 | 0.117 | 0.159 | 0.199 | 0.190 | 0.198 | 0.211 | 0.234 | 0.001 | 0.000 |
| 2011 | 0.047 | 0.104 | 0.162 | 0.171 | 0.192 | 0.196 | 0.199 | 0.211 | 0.000 | 0.000 |
| 2012 | 0.052 | 0.093 | 0.142 | 0.188 | 0.198 | 0.206 | 0.215 | 0.215 | 0.000 | 0.000 |
| 2013 | 0.051 | 0.081 | 0.127 | 0.151 | 0.170 | 0.194 | 0.228 | 0.346 | 0.000 | 0.000 |
| 2014 | 0.025 | 0.089 | 0.132 | 0.162 | 0.180 | 0.212 | 0.300 | 0.370 | 0.255 | 0.000 |
| 2015 | 0.026 | 0.078 | 0.122 | 0.149 | 0.164 | 0.185 | 0.173 | 0.218 | 0.404 | 0.291 |

Table 8.2.8 Plaice in 4 and 3.a. Catch weight at age (kg).

| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1957 | 0.044 | 0.110 | 0.194 | 0.257 | 0.349 | 0.455 | 0.533 | 0.589 | 0.396 | 0.998 |
| 1958 | 0.047 | 0.106 | 0.189 | 0.256 | 0.329 | 0.452 | 0.513 | 0.615 | 0.665 | 0.992 |
| 1959 | 0.051 | 0.120 | 0.192 | 0.260 | 0.345 | 0.472 | 0.592 | 0.623 | 0.750 | 1.000 |
| 1960 | 0.045 | 0.115 | 0.204 | 0.287 | 0.377 | 0.480 | 0.601 | 0.683 | 0.724 | 1.094 |
| 1961 | 0.044 | 0.101 | 0.180 | 0.301 | 0.402 | 0.506 | 0.604 | 0.671 | 0.812 | 1.071 |
| 1962 | 0.042 | 0.099 | 0.181 | 0.273 | 0.397 | 0.538 | 0.570 | 0.692 | 0.777 | 1.127 |
| 1963 | 0.048 | 0.110 | 0.175 | 0.304 | 0.392 | 0.531 | 0.624 | 0.667 | 0.715 | 1.028 |
| 1964 | 0.032 | 0.126 | 0.204 | 0.271 | 0.379 | 0.485 | 0.628 | 0.700 | 0.737 | 1.005 |
| 1965 | 0.038 | 0.076 | 0.214 | 0.313 | 0.381 | 0.469 | 0.539 | 0.663 | 0.726 | 0.887 |
| 1966 | 0.038 | 0.104 | 0.148 | 0.315 | 0.428 | 0.483 | 0.559 | 0.624 | 0.690 | 0.933 |
| 1967 | 0.036 | 0.111 | 0.190 | 0.235 | 0.422 | 0.543 | 0.597 | 0.662 | 0.738 | 0.978 |
| 1968 | 0.060 | 0.116 | 0.223 | 0.275 | 0.340 | 0.516 | 0.590 | 0.596 | 0.686 | 0.911 |
| 1969 | 0.052 | 0.174 | 0.272 | 0.284 | 0.356 | 0.408 | 0.573 | 0.655 | 0.658 | 0.893 |
| 1970 | 0.049 | 0.131 | 0.268 | 0.352 | 0.394 | 0.441 | 0.499 | 0.672 | 0.744 | 0.892 |
| 1971 | 0.057 | 0.160 | 0.277 | 0.388 | 0.444 | 0.512 | 0.542 | 0.607 | 0.699 | 0.891 |
| 1972 | 0.067 | 0.207 | 0.290 | 0.407 | 0.486 | 0.540 | 0.608 | 0.646 | 0.674 | 0.939 |
| 1973 | 0.045 | 0.205 | 0.334 | 0.403 | 0.478 | 0.538 | 0.605 | 0.627 | 0.677 | 0.842 |
| 1974 | 0.056 | 0.121 | 0.343 | 0.404 | 0.472 | 0.552 | 0.609 | 0.693 | 0.707 | 0.926 |
| 1975 | 0.069 | 0.152 | 0.205 | 0.393 | 0.492 | 0.585 | 0.636 | 0.703 | 0.783 | 1.019 |
| 1976 | 0.088 | 0.181 | 0.262 | 0.347 | 0.509 | 0.592 | 0.641 | 0.705 | 0.741 | 0.980 |
| 1977 | 0.071 | 0.218 | 0.245 | 0.318 | 0.396 | 0.551 | 0.647 | 0.721 | 0.715 | 0.978 |
| 1978 | 0.070 | 0.190 | 0.315 | 0.364 | 0.432 | 0.486 | 0.609 | 0.687 | 0.776 | 0.950 |
| 1979 | 0.067 | 0.190 | 0.295 | 0.338 | 0.426 | 0.472 | 0.550 | 0.675 | 0.796 | 0.960 |
| 1980 | 0.056 | 0.197 | 0.343 | 0.399 | 0.471 | 0.542 | 0.588 | 0.662 | 0.772 | 1.013 |
| 1981 | 0.048 | 0.183 | 0.327 | 0.415 | 0.502 | 0.556 | 0.604 | 0.642 | 0.725 | 1.007 |
| 1982 | 0.056 | 0.152 | 0.308 | 0.421 | 0.512 | 0.606 | 0.664 | 0.712 | 0.738 | 0.984 |
| 1983 | 0.052 | 0.153 | 0.275 | 0.379 | 0.507 | 0.602 | 0.677 | 0.771 | 0.815 | 0.984 |
| 1984 | 0.053 | 0.150 | 0.265 | 0.321 | 0.470 | 0.588 | 0.677 | 0.726 | 0.839 | 1.036 |
| 1985 | 0.054 | 0.169 | 0.268 | 0.333 | 0.444 | 0.562 | 0.667 | 0.730 | 0.807 | 1.021 |
| 1986 | 0.049 | 0.141 | 0.275 | 0.311 | 0.402 | 0.472 | 0.668 | 0.750 | 0.856 | 1.014 |
| 1987 | 0.043 | 0.113 | 0.219 | 0.343 | 0.375 | 0.471 | 0.574 | 0.728 | 0.835 | 0.993 |
| 1988 | 0.043 | 0.102 | 0.197 | 0.276 | 0.415 | 0.477 | 0.590 | 0.680 | 0.808 | 1.017 |
| 1989 | 0.047 | 0.118 | 0.215 | 0.291 | 0.364 | 0.512 | 0.591 | 0.668 | 0.785 | 0.940 |
| 1990 | 0.053 | 0.130 | 0.211 | 0.290 | 0.360 | 0.440 | 0.586 | 0.690 | 0.761 | 1.010 |
| 1991 | 0.056 | 0.149 | 0.211 | 0.273 | 0.349 | 0.449 | 0.526 | 0.666 | 0.743 | 0.924 |
| 1992 | 0.055 | 0.145 | 0.226 | 0.275 | 0.324 | 0.410 | 0.530 | 0.607 | 0.719 | 0.891 |
| 1993 | 0.063 | 0.160 | 0.250 | 0.303 | 0.340 | 0.407 | 0.512 | 0.630 | 0.720 | 0.856 |
| 1994 | 0.064 | 0.179 | 0.257 | 0.331 | 0.372 | 0.416 | 0.491 | 0.610 | 0.731 | 0.906 |
| 1995 | 0.071 | 0.183 | 0.283 | 0.334 | 0.373 | 0.419 | 0.474 | 0.593 | 0.734 | 0.906 |
| 1996 | 0.054 | 0.140 | 0.268 | 0.336 | 0.413 | 0.464 | 0.490 | 0.553 | 0.712 | 0.858 |
| 1997 | 0.045 | 0.129 | 0.220 | 0.353 | 0.408 | 0.473 | 0.541 | 0.574 | 0.616 | 0.912 |
| 1998 | 0.047 | 0.094 | 0.208 | 0.299 | 0.449 | 0.544 | 0.613 | 0.673 | 0.687 | 0.899 |
| 1999 | 0.054 | 0.103 | 0.199 | 0.267 | 0.413 | 0.414 | 0.538 | 0.637 | 0.748 | 0.804 |


|  | AGE |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ |
| 2000 | 0.063 | 0.123 | 0.210 | 0.274 | 0.372 | 0.452 | 0.565 | 0.601 | 0.752 | 0.888 |
| 2001 | 0.090 | 0.136 | 0.197 | 0.234 | 0.303 | 0.410 | 0.577 | 0.701 | 0.796 | 0.799 |
| 2002 | 0.057 | 0.131 | 0.222 | 0.285 | 0.326 | 0.426 | 0.469 | 0.644 | 0.760 | 0.904 |
| 2003 | 0.066 | 0.124 | 0.226 | 0.284 | 0.336 | 0.385 | 0.419 | 0.635 | 0.763 | 0.857 |
| 2004 | 0.054 | 0.125 | 0.220 | 0.297 | 0.376 | 0.421 | 0.506 | 0.560 | 0.797 | 0.872 |
| 2005 | 0.067 | 0.117 | 0.213 | 0.298 | 0.352 | 0.347 | 0.453 | 0.554 | 0.617 | 0.910 |
| 2006 | 0.060 | 0.139 | 0.212 | 0.293 | 0.375 | 0.383 | 0.428 | 0.457 | 0.531 | 0.748 |
| 2007 | 0.059 | 0.114 | 0.223 | 0.310 | 0.351 | 0.375 | 0.491 | 0.357 | 0.587 | 0.632 |
| 2008 | 0.057 | 0.123 | 0.248 | 0.325 | 0.389 | 0.437 | 0.368 | 0.469 | 0.640 | 0.638 |
| 2009 | 0.061 | 0.125 | 0.232 | 0.327 | 0.399 | 0.466 | 0.518 | 0.441 | 0.668 | 0.792 |
| 2010 | 0.062 | 0.132 | 0.226 | 0.311 | 0.396 | 0.442 | 0.463 | 0.574 | 0.682 | 0.649 |
| 2011 | 0.048 | 0.115 | 0.212 | 0.277 | 0.369 | 0.453 | 0.595 | 0.445 | 0.556 | 0.804 |
| 2012 | 0.052 | 0.096 | 0.196 | 0.294 | 0.348 | 0.425 | 0.509 | 0.557 | 0.558 | 0.680 |
| 2013 | 0.051 | 0.091 | 0.179 | 0.274 | 0.345 | 0.409 | 0.490 | 0.599 | 0.607 | 0.680 |
| 2014 | 0.025 | 0.093 | 0.172 | 0.253 | 0.318 | 0.396 | 0.473 | 0.542 | 0.628 | 0.650 |
| 2015 | 0.026 | 0.080 | 0.162 | 0.258 | 0.326 | 0.394 | 0.461 | 0.481 | 0.582 | 0.600 |

Table 8.2.9 North Sea plaice (including Skagerrak). Natural mortality at age and maturity ate age

| AGE | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| natural mortality | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| maturity | 0 | 0.5 | 0.5 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 8.2.10 North Sea plaice. Survey tuning indices

| BTS-ISIS | AGE |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| year | 1 | 2 | 3 | 4 | 6 | 7 | 8 | 9 |  |
| 1985 | 137 | 173.9 | 36.1 | 11 | 1.27 | 0.973 | 0.336 | 0.155 | 0.091 |
| 1986 | 667 | 131.7 | 50.2 | 9.21 | 3.78 | 0.4 | 0.418 | 0.147 | 0.07 |
| 1987 | 226 | 764.2 | 33.8 | 4.88 | 1.84 | 0.607 | 0.252 | 0.134 | 0.078 |
| 1988 | 680 | 147 | 182.3 | 9.99 | 2.81 | 0.814 | 0.458 | 0.036 | 0.112 |
| 1989 | 468 | 319.3 | 38.7 | 47.3 | 5.85 | 0.833 | 0.311 | 0.661 | 0.132 |
| 1990 | 185 | 146.1 | 79.3 | 26.35 | 5.47 | 0.758 | 0.189 | 0.383 | 0.239 |
| 1991 | 291 | 159.4 | 34 | 13.57 | 4.31 | 5.659 | 0.239 | 0.204 | 0.092 |
| 1992 | 361 | 174.5 | 29.3 | 5.96 | 3.75 | 2.871 | 1.186 | 0.346 | 0.05 |
| 1993 | 189 | 283.4 | 62.8 | 8.27 | 1.13 | 1.13 | 0.584 | 0.464 | 0.155 |
| 1994 | 193 | 77.1 | 34.5 | 10.59 | 2.67 | 0.6 | 0.8 | 0.895 | 0.373 |
| 1995 | 266 | 40.6 | 13.2 | 7.53 | 1.11 | 0.806 | 0.33 | 1.051 | 0.202 |


| BTS-COMbined | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 143.9 | 99.6 | 13.3 | 4.27 | 3.04 | 1.65 | 0.676 | 0.442 | 0.214 |
| 1997 | 386.8 | 28.7 | 14.9 | 4.01 | 2.04 | 1.54 | 0.428 | 0.797 | 0.327 |
| 1998 | 131.2 | 177.6 | 25.5 | 7.27 | 2.5 | 1.35 | 0.955 | 0.808 | 0.323 |
| 1999 | 117 | 53.6 | 96.3 | 6.49 | 3 | 1.93 | 0.659 | 0.756 | 0.314 |
| 2000 | 108.4 | 38.9 | 22.9 | 23.68 | 3.02 | 1.73 | 1.113 | 0.797 | 0.219 |
| 2001 | 80.3 | 39.8 | 15.7 | 8.75 | 9.3 | 1.08 | 0.624 | 0.42 | 0.511 |
| 2002 | 217.3 | 26.7 | 14 | 7.62 | 4.79 | 4.64 | 0.754 | 0.765 | 0.385 |
| 2003 | 53.6 | 94.4 | 15.9 | 10.3 | 5.36 | 3.08 | 4.007 | 0.732 | 0.76 |
| 2004 | 101.4 | 30.3 | 51.2 | 11.21 | 4.96 | 2.88 | 1.538 | 3.402 | 0.391 |
| 2005 | 70.8 | 45.6 | 13.8 | 20.39 | 3.04 | 6.94 | 1.568 | 0.571 | 3.57 |
| 2006 | 54.9 | 42.9 | 29.2 | 11.75 | 12.05 | 2.11 | 3.938 | 0.844 | 0.767 |
| 2007 | 139.4 | 44.4 | 24.6 | 26.58 | 5.68 | 11.69 | 2.091 | 3.947 | 0.364 |
| 2008 | 98.9 | 89.7 | 33.8 | 20.73 | 20.61 | 6.33 | 13.054 | 2.727 | 6.718 |
| 2009 | 170.8 | 76.5 | 54.1 | 21.48 | 12.83 | 12.19 | 3.139 | 10.254 | 1.585 |
| 2010 | 144.8 | 69.5 | 47.9 | 40.35 | 17.91 | 6.84 | 15.841 | 3.179 | 8.306 |
| 2011 | 226.5 | 126 | 58.1 | 32.75 | 33.17 | 15.09 | 5.808 | 11.94 | 1.124 |
| 2012 | 118.4 | 149.6 | 79.8 | 35.86 | 22.17 | 16.39 | 7.216 | 3.544 | 8.696 |
| 2013 | 192.8 | 90.5 | 90.3 | 46.71 | 27.6 | 15.37 | 11.273 | 4.523 | 3.224 |
| 2014 | 155.2 | 123.2 | 83.3 | 58.53 | 34.74 | 14.87 | 10.569 | 6.607 | 7.591 |
| 2015 | 116.5 | 156.6 | 102.5 | 57.4 | 49.2 | 25.5 | 9.7 | 7.0 | 7.4 |


| SNS1 |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | age |  |  |  |  |  | age |  |  |  |  |
| year | 1 | 2 | 3 | 4 | 5 | year | 1 | 2 | 3 | 4 | 5 |
| 1970 | 9311 | 9732 | 3273 | 770 | 170 | 2000 | 22855 | 2493 | 891 | 983 | 17 |
| 1971 | 13538 | 28164 | 1415 | 101 | 50 | 2001 | 11511 | 2898 | 370 | 176 | 691 |
| 1972 | 13207 | 10780 | 4478 | 89 | 84 | 2002 | 30809 | 1103 | 265 | 65 | 69 |
| 1973 | 65643 | 5133 | 1578 | 461 | 15 | 2003 | NA | NA | NA | NA | NA |
| 1974 | 15366 | 16509 | 1129 | 160 | 82 | 2004 | 18202 | 1350 | 1081 | 51 | 27 |
| 1975 | 11628 | 8168 | 9556 | 65 | 15 | 2005 | 10118 | 1819 | 142 | 366 | 8 |
| 1976 | 8537 | 2403 | 868 | 236 | 0 | 2006 | 12164 | 1571 | 385 | 52 | 54 |
| 1977 | 18537 | 3424 | 1737 | 590 | 213 | 2007 | 14175 | 2134 | 140 | 52 | 0 |
| 1978 | 14012 | 12678 | 345 | 135 | 45 | 2008 | 14706 | 2700 | 464 | 179 | 34 |
| 1979 | 21495 | 9829 | 1575 | 161 | 17 | 2009 | 14860 | 2019 | 492 | 38 | 20 |
| 1980 | 59174 | 12882 | 491 | 180 | 24 | 2010 | 11947 | 1812 | 529 | 55 | 10 |
| 1981 | 24756 | 18785 | 834 | 38 | 32 | 2011 | 18349 | 1143 | 308 | 75 | 60 |
| 1982 | 69993 | 8642 | 1261 | 88 | 8 | 2012 | 5893 | 2929 | 682 | 82 | 30 |
| 1983 | 33974 | 13909 | 249 | 71 | 6 | 2013 | 15395 | 3021 | 1638 | 428 | 89 |
| 1984 | 44965 | 10413 | 2467 | 42 | 0 | 2014 | 17313 | 2258 | 514 | 458 | 58 |
| 1985 | 28101 | 13848 | 1598 | 328 | 17 | 2015 | 16727 | 5040 | 1882 | 478 | 200 |
| 1986 | 93552 | 7580 | 1152 | 145 | 30 |  |  |  |  |  |  |
| 1987 | 33402 | 32991 | 1227 | 200 | 30 |  |  |  |  |  |  |
| 1988 | 36609 | 14421 | 13153 | 1350 | 88 |  |  |  |  |  |  |
| 1989 | 34276 | 17810 | 4373 | 7126 | 289 |  |  |  |  |  |  |
| 1990 | 25037 | 7496 | 3160 | 816 | 422 |  |  |  |  |  |  |
| 1991 | 57221 | 11247 | 1518 | 1077 | 128 |  |  |  |  |  |  |
| 1992 | 46798 | 13842 | 2268 | 613 | 176 |  |  |  |  |  |  |
| 1993 | 22098 | 9686 | 1006 | 98 | 60 |  |  |  |  |  |  |
| 1994 | 19188 | 4977 | 856 | 76 | 23 |  |  |  |  |  |  |
| 1995 | 24767 | 2796 | 381 | 97 | 38 |  |  |  |  |  |  |
| 1996 | 23015 | 10268 | 1185 | 45 | 47 |  |  |  |  |  |  |
| 1997 | 95901 | 4473 | 497 | 32 | 0 |  |  |  |  |  |  |
| 1998 | 33666 | 30242 | 5014 | 50 | 10 |  |  |  |  |  |  |
| 1999 | 32951 | 10272 | 13783 | 1058 | 17 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |

Table 8.3.1 North Sea plaice (4, 3.a and 7.dQ1). XSA diagnostics from final run
FLR XSA Diagnostics 2016-04-28 10:15:08
CPUE data from indices
Catch data for 59 years. 1957 to 2015. Ages 1 to 10.
fleet first age last age first year last year alpha beta
1 BTS-Isis-early $1 \quad 8 \quad 198519950.660 .75$
$\begin{array}{llllllllllll}\text { BTS-Combined } & 1 & 9 & 1996 & 2015 & 0.66 & 0.75\end{array}$
$\begin{array}{llllllll}\text { SNS1 } & 1 & 3 & 1982 & 1999 & 0.66 & 0.75\end{array}$
$\begin{array}{lllllllll}\text { SNS2 } & 1 & 3 & 2000 & 2015 & 0.66 & 0.75\end{array}$
Time series weights :
Tapered time weighting not applied
Catchability analysis :
Catchability independent of size for all ages
Catchability independent of age for ages >= 6
Terminal population estimation :
Survivor estimates shrunk towards the mean F of the final 5 years or the 5 oldest ages.
S.E. of the mean to which the estimates are shrunk $=2$

Minimum standard error for population
estimates derived from each fleet $=0.3$
prior weighting not applied
Regression weights
year
age 2006200720082009201020112012201320142015
all $1 \begin{array}{llllllllll} & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$
Fishing mortalities
year
age 2006200720082009201020112012201320142015
10.2710 .0750 .1430 .1530 .1340 .0680 .0790 .0830 .1710 .080
20.5260 .4470 .3490 .2970 .2800 .1700 .2350 .1800 .1590 .230
$30.4330 .440 \quad 0.2790 .2540 .227 \quad 0.2140 .2560 .1910 .2290 .182$
40.4200 .2380 .3140 .2290 .2570 .2950 .2640 .1960 .1790 .213
50.2420 .3400 .1650 .2430 .1650 .2560 .2340 .1800 .1840 .127
60.2390 .1780 .1760 .1150 .1770 .1350 .1830 .1780 .1210 .115
70.1090 .1250 .1350 .1000 .0850 .1070 .0730 .1090 .1000 .080
80.2980 .0760 .0870 .0730 .0730 .0800 .0650 .0430 .0710 .082
90.1630 .0970 .0180 .0320 .0290 .0500 .0350 .0550 .0310 .040
100.1630 .0970 .0180 .0320 .0290 .0500 .0350 .0550 .0310 .040

XSA population number (Thousand)

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 979159 | 635267 | 590473 | 134811 | 261313 | 41158 | 64566 | 10197 | 9220 | 9832 |  |  |  |
| 2007 | 1143879 | 675430 | 339709 | 346628 | 80123 | 185625 | 29338 | 52405 | 6852 | 14779 |  |  |  |
| 2008 | 1071558 | 959761 | 390694 | 198006 | 247160 | 51596 | 140503 | 23421 | 43941 | 55932 |  |  |  |
| 2009 | 1110840 | 840310 | 612415 | 267556 | 130943 | 189566 | 39169 | 111040 | 19420 | 27170 |  |  |  |
| 2010 | 1419551 | 862227 | 564777 | 429637 | 192616 | 92901 | 152833 | 32077 | 93358 | 46909 |  |  |  |
| 2011 | 1892434 | 1123228 | 589703 | 407373 | 300783 | 147766 | 70403 | 127027 | 26980 | 58621 |  |  |  |
| 2012 | 1274853 | 1599807 | 857721 | 430806 | 274371 | 210777 | 116830 | 57248 | 106134 | 58598 |  |  |  |
| 2013 | 1703575 | 1065903 | 1143954 | 600668 | 299390 | 196508 | 158773 | 98310 | 48558 | 74607 |  |  |  |
| 2014 | 1966051 | 1418962 | 805264 | 855078 | 446622 | 226197 | 148777 | 128855 | 85178 | 118469 |  |  |  |
| 2015 | 1140208 | 1499750 | 1095662 | 579455 | 646926 | 336195 | 181327 | 121842 | 108551 | 129321 |  |  |  |

Estimated population abundance at 1st Jan 2016
age
$\begin{array}{lllllllllll}\text { year } & 1 & 2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10\end{array}$
20160952367107852482640342358851543527124715142510159594378
Fleet: BTS-Isis-early
Log catchability residuals.
year
age $19851986198719881989 \quad 1990 \quad 1991 \quad 1992 \quad 199319941995$
1-1.127-0.461-0.707 $0.481 \quad 0.470-0.344 \quad 0.2810 .647 \quad 0.362 \quad 0.507-0.110$
$20.027-0.553$ 0.323 -0.528 0.297-0.206 0.067 0.314 0.904-0.045 -0.601 3-0.269 0. $212-0.397 \quad 0.415-0.390 \quad 0.329-0.238-0.171 \quad 0.7020 .213-0.407$ 4-0.366-0.226-0.600-0.091 0.542 0.652 0.023-0.531 0.007 0.374 0. 216

```
5-0.495 0.063 -0.292 0.405 0.828 -0.163 0.232 0.279 -0.743 0.270 -0.384
6 0.293-0.710 -0.746 -0.027 0.186 -0.278 0.856 0.714 0.127 -0.424 0.008
7 0.056 0.096 -0.282 -0.212 -0.207 -0.601 -0.649 -0.015 -0.314 0.719 -0.201
8-0.102 -0.074 -0.423 -1.162 0.918 0.654 0.235 0.559 -0.435 0.544 1.811
Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time
```



```
S.E_Logq 0.5047 0.5047 0.5047 0.5047 0.5047 0.5047 0.5047 0.5047
```

Fleet: BTS-Combined
Log catchability residuals.
year
age 199619971998199920002001200220032004200520062007
1 -0.02 0.44 0.42 0.24-0.04 0.22 0.13 -0.15-0.33-0.19 -0.55 0.09
$20.44-0.790 .31-0.17-0.28-0.24-0.19 \quad 0.050 .04-0.46 \quad 0.01-0.07$
3-0.01-0.54 0.11 0.36-0.17 -0.29 -0.58 0.02 0.09-0.10-0.30 0.09
$4-0.46-0.41-0.15-0.13-0.33-0.27-0.13-0.25 \quad 0.39-0.27-0.34-0.09$
$5-0.23-0.29-0.14-0.13-0.03-0.41 \quad 0.01 \quad 0.36-0.53-0.31-0.370 .13$
$6-0.27-0.07-0.05 \quad 0.38-0.04-0.43-0.50 \quad 0.32 \quad 0.41 \quad 0.09-0.32-0.15$
$\begin{array}{lllllllllllllllllll}7 & -0.64 & -0.82 & 0.18 & -0.25 & 0.41 & -0.54 & -0.10 & -0.11 & 0.15 & 0.38 & -0.23 & -0.06\end{array}$
8 -0.37 0.29 0.44 0.54 0.28-0.23 0.05 0.36 0.04-0.47 0.21-0.04
$9-0.210 .210 .010 .02-0.27 \quad 0.08-0.04 \quad 0.29-0.020 .270 .12-0.38$
year
age 20082009201020112012201320142015
$1-0.14-0.38-0.05 \quad 0.07-0.18 \quad 0.02-0.28-0.08$
$20.210 .150 .020 .270 .13-0.010 .000 .24$
$30.150 .160 .100 .240 .21-0.000 .300 .16$
$\begin{array}{lllllllllllll} & 0.45 & 0.13 & 0.30 & 0.17 & 0.19 & 0.07 & -0.07 & 0.33\end{array}$
50.160 .380 .270 .510 .180 .280 .110 .04
6 0.51-0.17 $0.01 \quad 0.30 \quad 0.06 \quad 0.07-0.15-0.01$
$7 \quad 0.21 \quad 0.04 \quad 0.28 \quad 0.07-0.24-0.08-0.09-0.38$
80.40 0.19 $0.23-0.18-0.25-0.56-0.43-0.31$
$9 \quad 0.62 \quad 0.01 \quad 0.09-0.65 \quad 0.01-0.19 \quad 0.09-0.18$
Mean $\log$ catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

```
    1 2 3 4 4 5 6 % 7 8 9
Mean_Logq -8.979 -9.172 -9.241 -9.324 -9.369 -9.326 -9.326 -9.326 -9.326
S.E_Logq 0.284 0.284 0.284 0.284 0.284 0.284 0.284 0.284 0.284
```

Fleet: SNS1
Log catchability residuals.
year
age 198219831984198519861987198819891990199119921993
1 0.16-0.14 0.24-0.62-0.34 -0.53-0.35-0.06 -0.26 0.74 0.69 0.30
$20.12-0.20-0.04 \quad 0.31-0.59-0.00-0.03 \quad 0.23-0.36 \quad 0.230 .60 \quad 0.34$

year
age $19941995 \quad 1996 \quad 199719981999$
$10.286-0.394-0.5970 .308 \quad 0.3190 .230$
$20.030-0.461-0.022-0.833 \quad 0.3560 .335$
3-0.293-0.765 0.474-1.034 1.391 1.314
Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

```
    1 2 3
Mean_Logq -3.3312 -4.0788 -5.2354
S.E_Logq 0.5496 0.5496 0.5496
```

Fleet: SNS2
Log catchability residuals.
year
age 2000200120022003200420052006200720082009201020112012
10.60 0.46 0.37 NA $0.150 .05 \quad 0.14-0.01 \quad 0.14 \quad 0.13-0.35-0.26-0.99$
20.420 .590 .07 NA $0.37-0.240 .15$ 0.34 0.15 -0.04 -0.19-0.99 -0. 36
$30.95-0.33-0.16$ NA $0.60-0.32-0.27-0.72 \quad 0.23-0.18-0.05-0.64-0.19$
year
age 201320142015
$1-0.32-0.28 \quad 0.17$
$20.04-0.550 .25$
$30.35-0.430 .53$
Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

```
    1 2 3
Mean_Logq -4.2612 -5.7091 -6.6948
S.E_Logq 0.4200 0.4200 0.4200
```

Terminal year survivor and $F$ summaries:
Age 1 Year class =2014
source scaledWts survivors yrcls
BTS-Combined 0.637 8761872014
SNS2 0.348 11245452014
fshk $\quad 0.016 \quad 7023042014$
Age 2 Year class =2013
source scaledWts survivors yrcls BTS-Combined 0.65713679912013
SNS2 $\quad 0.325 \quad 13790922013$
fshk 0.019 12230622013
Age 3 Year class $=2012$
source scaledWts survivors yrcls
BTS-Combined $0.717 \quad 9719702012$
SNS2 0.264 13995192012
fshk 0.019 6577842012
Age 4 Year class =2011
source scaledWts survivors yrcls BTS-Combined 0.9735864942011 fshk 0.027 3737132011

Age 5 Year class $=2010$
source scaledWts survivors yrcls BTS-Combined $0.975 \quad 5387072010$ fshk 0.025 3089472010

Age 6 Year class $=2009$
source scaledWts survivors yrcls
BTS-Combined 0.9752687652009
fshk 0.0251910932009
Age 7 Year class $=2008$
source scaledWts survivors yrcls
BTS-Combined 0.9681033952008
fshk $0.032 \quad 1275242008$
Age 8 Year class =2007
source scaledWts survivors yrcls
BTS-Combined 0.969 747892007
fshk $0.031 \quad 1260272007$
Age 9 Year class =2006
source scaledWts survivors yrcls BTS-Combined 0.977 792192006
fshk 0.023 292882006

Table 8.3.2 Plaice in 4 and 3.a. Harvest (F)

| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1957 | 0.077 | 0.228 | 0.271 | 0.325 | 0.370 | 0.224 | 0.293 | 0.336 | 0.311 | 0.311 |
| 1958 | 0.105 | 0.250 | 0.308 | 0.373 | 0.389 | 0.336 | 0.280 | 0.303 | 0.337 | 0.337 |
| 1959 | 0.152 | 0.31 | 0.356 | 0.379 | 0.416 | 0.389 | 0.355 | 0.314 | 0.372 | 0.372 |
| 1960 | 0.108 | 0.317 | 0.353 | 0.385 | 0.366 | 0.419 | 0.383 | 0.359 | 0.384 | 0.384 |
| 1961 | 0.096 | 0.288 | 0.345 | 0.362 | 0.423 | 0.334 | 0.365 | 0.443 | 0.386 | 0.386 |
| 1962 | 0.095 | 0.316 | 0.366 | 0.385 | 0.417 | 0.409 | 0.347 | 0.336 | 0.380 | 0.380 |
| 1963 | 0.148 | 0.361 | 0.415 | 0.435 | 0.425 | 0.476 | 0.452 | 0.454 | 0.450 | 0.450 |
| 1964 | 0.031 | 0.394 | 0.443 | 0.464 | 0.536 | 0.484 | 0.400 | 0.387 | 0.455 | 0.455 |
| 1965 | 0.068 | 0.264 | 0.391 | 0.407 | 0.350 | 0.501 | 0.411 | 0.347 | 0.404 | 0.404 |
| 1966 | 0.071 | 0.352 | 0.382 | 0.426 | 0.475 | 0.341 | 0.503 | 0.406 | 0.432 | 0.432 |
| 1967 | 0.054 | 0.347 | 0.400 | 0.403 | 0.472 | 0.501 | 0.307 | 0.432 | 0.425 | 0.425 |
| 1968 | 0.195 | 0.283 | 0.342 | 0.359 | 0.366 | 0.323 | 0.410 | 0.289 | 0.350 | 0.350 |
| 1969 | 0.146 | 0.312 | 0.336 | 0.348 | 0.327 | 0.445 | 0.304 | 0.414 | 0.369 | 0.369 |
| 1970 | 0.218 | 0.424 | 0.473 | 0.482 | 0.441 | 0.481 | 0.567 | 0.247 | 0.445 | 0.445 |
| 1971 | 0.191 | 0.325 | 0.383 | 0.387 | 0.407 | 0.394 | 0.427 | 0.411 | 0.406 | 0.406 |
| 1972 | 0.227 | 0.373 | 0.394 | 0.409 | 0.416 | 0.44 | 0.404 | 0.474 | 0.43 | 0.43 |
| 1973 | 0.113 | 0.391 | 0.438 | 0.548 | 0.553 | 0.419 | 0.393 | 0.487 | 0.482 | 0.482 |
| 1974 | 0.220 | 0.396 | 0.494 | 0.517 | 0.598 | 0.455 | 0.397 | 0.467 | 0.489 | 0.489 |
| 1975 | 0.355 | 0.503 | 0.535 | 0.577 | 0.624 | 0.642 | 0.498 | 0.524 | 0.575 | 0.575 |
| 1976 | 0.334 | 0.410 | 0.432 | 0.441 | 0.393 | 0.444 | 0.53 | 0.463 | 0.456 | 0.456 |
| 1977 | 0.323 | 0.473 | 0.497 | 0.503 | 0.670 | 0.423 | 0.444 | 0.537 | 0.517 | 0.517 |
| 1978 | 0.302 | 0.422 | 0.452 | 0.456 | 0.445 | 0.501 | 0.445 | 0.411 | 0.453 | 0.453 |
| 1979 | 0.418 | 0.627 | 0.655 | 0.666 | 0.672 | 0.696 | 0.693 | 0.596 | 0.667 | 0.667 |
| 1980 | 0.233 | 0.459 | 0.662 | 0.621 | 0.511 | 0.515 | 0.491 | 0.501 | 0.530 | 0.530 |
| 1981 | 0.170 | 0.476 | 0.572 | 0.604 | 0.585 | 0.456 | 0.506 | 0.536 | 0.539 | 0.539 |
| 1982 | 0.231 | 0.489 | 0.678 | 0.659 | 0.597 | 0.519 | 0.485 | 0.531 | 0.560 | 0.560 |
| 1983 | 0.225 | 0.487 | 0.514 | 0.709 | 0.568 | 0.510 | 0.450 | 0.476 | 0.545 | 0.545 |
| 1984 | 0.290 | 0.513 | 0.537 | 0.588 | 0.649 | 0.563 | 0.489 | 0.556 | 0.571 | 0.571 |
| 1985 | 0.257 | 0.447 | 0.443 | 0.656 | 0.488 | 0.518 | 0.513 | 0.485 | 0.534 | 0.534 |
| 1986 | 0.281 | 0.590 | 0.585 | 0.596 | 0.730 | 0.718 | 0.560 | 0.632 | 0.729 | 0.729 |
| 1987 | 0.213 | 0.632 | 0.647 | 0.694 | 0.821 | 0.726 | 0.819 | 0.501 | 0.664 | 0.664 |
| 1988 | 0.223 | 0.603 | 0.646 | 0.678 | 0.800 | 0.775 | 0.711 | 1.028 | 0.752 | 0.752 |
| 1989 | 0.199 | 0.549 | 0.584 | 0.608 | 0.660 | 0.664 | 0.643 | 0.644 | 1.356 | 1.356 |
| 1990 | 0.154 | 0.438 | 0.537 | 0.563 | 0.689 | 0.638 | 0.567 | 0.515 | 0.642 | 0.642 |
| 1991 | 0.226 | 0.563 | 0.587 | 0.631 | 0.617 | 0.693 | 0.703 | 0.698 | 0.715 | 0.715 |
| 1992 | 0.204 | 0.510 | 0.578 | 0.592 | 0.806 | 0.587 | 0.549 | 0.752 | 1.145 | 1.145 |
| 1993 | 0.205 | 0.453 | 0.533 | 0.586 | 0.694 | 0.733 | 0.414 | 0.364 | 1.008 | 1.008 |
| 1994 | 0.149 | 0.437 | 0.565 | 0.622 | 0.615 | 0.643 | 0.882 | 0.296 | 0.261 | 0.261 |
| 1995 | 0.118 | 0.408 | 0.553 | 0.710 | 0.746 | 0.653 | 0.752 | 0.650 | 0.154 | 0.154 |
| 1996 | 0.092 | 0.525 | 0.591 | 0.638 | 0.700 | 0.665 | 0.715 | 0.682 | 0.769 | 0.769 |
| 1997 | 0.064 | 0.749 | 0.858 | 0.626 | 0.811 | 0.816 | 0.725 | 0.633 | 0.823 | 0.823 |
| 1998 | 0.145 | 0.476 | 0.896 | 0.954 | 0.587 | 0.504 | 0.550 | 0.438 | 0.485 | 0.485 |


| AGE |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1999 | 0.165 | 0.444 | 0.491 | 0.955 | 0.671 | 0.777 | 0.383 | 0.467 | 0.412 | 0.412 |
| 2000 | 0.114 | 0.343 | 0.303 | 0.556 | 0.634 | 0.550 | 0.354 | 0.182 | 0.273 | 0.273 |
| 2001 | 0.066 | 0.673 | 0.900 | 0.825 | 0.776 | 0.544 | 0.359 | 0.193 | 0.117 | 0.117 |
| 2002 | 0.194 | 0.541 | 0.473 | 0.609 | 0.704 | 0.453 | 0.607 | 0.256 | 0.167 | 0.167 |
| 2003 | 0.133 | 0.565 | 0.575 | 0.450 | 0.737 | 0.688 | 0.424 | 0.284 | 0.103 | 0.103 |
| 2004 | 0.197 | 0.594 | 0.453 | 0.508 | 0.243 | 0.552 | 0.320 | 0.124 | 0.099 | 0.099 |
| 2005 | 0.135 | 0.448 | 0.430 | 0.349 | 0.451 | 0.222 | 0.456 | 0.251 | 0.073 | 0.073 |
| 2006 | 0.271 | 0.526 | 0.433 | 0.420 | 0.242 | 0.239 | 0.109 | 0.298 | 0.163 | 0.163 |
| 2007 | 0.075 | 0.447 | 0.440 | 0.238 | 0.340 | 0.178 | 0.125 | 0.076 | 0.097 | 0.097 |
| 2008 | 0.143 | 0.349 | 0.279 | 0.314 | 0.165 | 0.176 | 0.135 | 0.087 | 0.018 | 0.018 |
| 2009 | 0.153 | 0.297 | 0.254 | 0.229 | 0.243 | 0.115 | 0.100 | 0.073 | 0.032 | 0.032 |
| 2010 | 0.134 | 0.280 | 0.227 | 0.257 | 0.165 | 0.177 | 0.085 | 0.073 | 0.029 | 0.029 |
| 2011 | 0.068 | 0.170 | 0.214 | 0.295 | 0.256 | 0.135 | 0.107 | 0.080 | 0.050 | 0.050 |
| 2012 | 0.079 | 0.235 | 0.256 | 0.264 | 0.234 | 0.183 | 0.073 | 0.065 | 0.035 | 0.035 |
| 2013 | 0.083 | 0.180 | 0.191 | 0.196 | 0.180 | 0.178 | 0.109 | 0.043 | 0.055 | 0.055 |
| 2014 | 0.171 | 0.159 | 0.229 | 0.179 | 0.184 | 0.121 | 0.100 | 0.071 | 0.031 | 0.031 |
| 2015 | 0.080 | 0.230 | 0.182 | 0.213 | 0.127 | 0.115 | 0.080 | 0.082 | 0.04 | 0.040 |
|  |  |  |  |  |  |  |  |  |  |  |

Table 8.3.3 Plaice in 4 and 3.a. Stock numbers (thousands)

|  | AGE |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1957 | 460518 | 260189 | 335426 | 191663 | 123254 | 51719 | 50633 | 36912 | 21748 | 47346 |
| 1958 | 700350 | 385916 | 187499 | 231548 | 125322 | 77050 | 37388 | 34164 | 23863 | 51185 |
| 1959 | 864891 | 570733 | 272015 | 124694 | 144305 | 76859 | 49836 | 25558 | 22821 | 55784 |
| 1960 | 760716 | 672161 | 378792 | 172422 | 77206 | 86127 | 47154 | 31612 | 16898 | 50156 |
| 1961 | 866067 | 617992 | 442786 | 240742 | 106211 | 48429 | 51269 | 29084 | 19973 | 48657 |
| 1962 | 593498 | 711758 | 419439 | 283718 | 151719 | 62981 | 31371 | 32208 | 16899 | 41003 |
| 1963 | 694671 | 488254 | 469573 | 263211 | 174662 | 90494 | 37844 | 20055 | 20836 | 48855 |
| 1964 | 2254825 | 542085 | 308041 | 280620 | 154097 | 103320 | 50871 | 21785 | 11522 | 48678 |
| 1965 | 701920 | 1977435 | 330605 | 178901 | 159702 | 81552 | 57635 | 30870 | 13392 | 55694 |
| 1966 | 594050 | 593547 | 1374584 | 202403 | 107760 | 101849 | 44696 | 34587 | 19750 | 45407 |
| 1967 | 407196 | 500900 | 377880 | 848488 | 119574 | 60657 | 65528 | 24461 | 20850 | 34495 |
| 1968 | 438895 | 349233 | 320398 | 229259 | 513144 | 67457 | 33243 | 43601 | 14369 | 48764 |
| 1969 | 658811 | 326766 | 238053 | 205988 | 144870 | 321955 | 44204 | 19957 | 29562 | 57869 |
| 1970 | 664222 | 515077 | 216461 | 153983 | 131672 | 94527 | 186702 | 29519 | 11931 | 42135 |
| 1971 | 420331 | 483400 | 304989 | 122040 | 86061 | 76619 | 52872 | 95822 | 20860 | 54789 |
| 1972 | 374300 | 314358 | 316083 | 188234 | 74988 | 51860 | 46745 | 31221 | 57490 | 48221 |
| 1973 | 1320356 | 269909 | 195949 | 192919 | 113137 | 44774 | 30229 | 28231 | 17585 | 69002 |
| 1974 | 1136000 | 1067543 | 165123 | 114422 | 100895 | 58912 | 26640 | 18472 | 15702 | 61709 |
| 1975 | 864714 | 824861 | 649780 | 91149 | 61716 | 50201 | 33806 | 16209 | 10480 | 50010 |
| 1976 | 691691 | 548413 | 451477 | 344430 | 46311 | 29915 | 23896 | 18591 | 8681 | 36821 |
| 1977 | 990829 | 448206 | 329361 | 265187 | 200493 | 28299 | 17369 | 12726 | 10588 | 23851 |
| 1978 | 920713 | 649359 | 252636 | 181325 | 145140 | 92799 | 16773 | 10083 | 6730 | 20692 |
| 1979 | 905430 | 616240 | 385404 | 145485 | 103978 | 84190 | 50856 | 9729 | 6049 | 19898 |
| 1980 | 1148883 | 539147 | 297895 | 181084 | 67660 | 48049 | 37964 | 23009 | 4851 | 13982 |
| 1981 | 901574 | 823274 | 308386 | 139032 | 88018 | 36735 | 25968 | 21017 | 12619 | 15283 |
| 1982 | 2111275 | 687932 | 462799 | 157493 | 68787 | 44350 | 21059 | 14169 | 11126 | 21400 |
| 1983 | 1368338 | 1515904 | 381875 | 212521 | 73755 | 34268 | 23889 | 11726 | 7541 | 20691 |
| 1984 | 1299663 | 988318 | 842738 | 206610 | 94598 | 37823 | 18611 | 13779 | 6593 | 15745 |
| 1985 | 1880989 | 880358 | 535656 | 445656 | 103845 | 44708 | 19494 | 10332 | 7148 | 14743 |
| 1986 | 4797263 | 1316261 | 509296 | 311173 | 209338 | 57683 | 24107 | 10560 | 5753 | 14286 |
| 1987 | 1979144 | 3276305 | 660311 | 256795 | 155113 | 91312 | 25449 | 12454 | 5081 | 12078 |
| 1988 | 1830953 | 1446915 | 1576342 | 312691 | 116109 | 61731 | 39972 | 10148 | 6828 | 12646 |
| 1989 | 1250820 | 1325118 | 716313 | 747916 | 143566 | 47195 | 25731 | 17765 | 3286 | 8129 |
| 1990 | 1084035 | 927707 | 692339 | 361401 | 368543 | 67131 | 21984 | 12244 | 8440 | 9281 |
| 1991 | 959797 | 841178 | 541628 | 366321 | 186293 | 167504 | 32089 | 11281 | 6618 | 11527 |
| 1992 | 811532 | 692911 | 433298 | 272479 | 176437 | 90942 | 75783 | 14377 | 5080 | 10650 |
| 1993 | 565366 | 599095 | 376429 | 219950 | 136380 | 71334 | 45739 | 39612 | 6133 | 9656 |
| 1994 | 480910 | 416644 | 344572 | 199927 | 110724 | 61658 | 31007 | 27361 | 24903 | 24879 |
| 1995 | 1197928 | 374746 | 243579 | 177120 | 97165 | 54149 | 29324 | 11620 | 18421 | 28716 |
| 1996 | 1339279 | 963465 | 225472 | 126796 | 78804 | 41713 | 25508 | 12506 | 5486 | 9713 |
| 1997 | 2212118 | 1104956 | 515628 | 112925 | 60614 | 35409 | 19408 | 11286 | 5721 | 8185 |
| 1998 | 813659 | 1878375 | 472787 | 197823 | 54665 | 24386 | 14175 | 8508 | 5420 | 9144 |


| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1999 | 882903 | 636602 | 1055790 | 174700 | 68947 | 27499 | 13329 | 7401 | 4968 | 10022 |
| 2000 | 1035754 | 677250 | 369519 | 584475 | 60803 | 31893 | 11436 | 8220 | 4199 | 9318 |
| 2001 | 577074 | 836291 | 435045 | 246925 | 303446 | 29177 | 16656 | 7265 | 6199 | 22518 |
| 2002 | 1857519 | 488990 | 386159 | 160044 | 97923 | 126368 | 15324 | 10521 | 5419 | 11151 |
| 2003 | 579018 | 1384282 | 257662 | 217666 | 78750 | 43823 | 72693 | 7553 | 7373 | 9401 |
| 2004 | 1375294 | 458760 | 712121 | 131183 | 125565 | 34113 | 19934 | 43056 | 5146 | 6360 |
| 2005 | 803930 | 1021614 | 229084 | 409430 | 71420 | 89130 | 17779 | 13097 | 34417 | 9658 |
| 2006 | 979159 | 635267 | 590473 | 134811 | 261313 | 41158 | 64566 | 10197 | 9220 | 9832 |
| 2007 | 1143879 | 675430 | 339709 | 346628 | 80123 | 185625 | 29338 | 52405 | 6852 | 14779 |
| 2008 | 1071558 | 959761 | 390694 | 198006 | 247160 | 51596 | 140503 | 23421 | 43941 | 55932 |
| 2009 | 1110840 | 840310 | 612415 | 267556 | 130943 | 189566 | 39169 | 111040 | 19420 | 27170 |
| 2010 | 1419551 | 862227 | 564777 | 429637 | 192616 | 92901 | 152833 | 32077 | 93358 | 46909 |
| 2011 | 1892434 | 1123228 | 589703 | 407373 | 300783 | 147766 | 70403 | 127027 | 26980 | 58621 |
| 2012 | 1274853 | 1599807 | 857721 | 430806 | 274371 | 210777 | 116830 | 57248 | 106134 | 58598 |
| 2013 | 1703575 | 1065903 | 1143954 | 600668 | 299390 | 196508 | 158773 | 98310 | 48558 | 74607 |
| 2014 | 1966051 | 1418962 | 805264 | 855078 | 446622 | 226197 | 148777 | 128855 | 85178 | 118469 |
| 2015 | 1140208 | 1499750 | 1095662 | 579455 | 646926 | 336195 | 181327 | 121842 | 108551 | 129321 |

Table 8.3.4 Plaice in 4 and 3.a. Stock summary table.

| YEAR | RECRUITS | SSB | CATCH | LANDINGS | DISCARDS | FBAR2-6 | FBAR HC2-6 | FBAR DIS2-3 | Y/SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1957 | 460518 | 274522 | 78443 | 70563 | 7880 | 0.28 | 0.23 | 0.12 | 0.26 |
| 1958 | 700350 | 285276 | 88191 | 73354 | 14837 | 0.33 | 0.25 | 0.19 | 0.26 |
| 1959 | 864891 | 290983 | 109164 | 79300 | 29864 | 0.37 | 0.24 | 0.24 | 0.27 |
| 1960 | 760716 | 300102 | 117334 | 87541 | 29793 | 0.37 | 0.27 | 0.23 | 0.29 |
| 1961 | 866067 | 313758 | 118474 | 85984 | 32490 | 0.35 | 0.24 | 0.27 | 0.27 |
| 1962 | 593498 | 373171 | 125375 | 87472 | 37903 | 0.38 | 0.24 | 0.29 | 0.23 |
| 1963 | 694671 | 359434 | 148376 | 107118 | 41258 | 0.42 | 0.27 | 0.35 | 0.30 |
| 1964 | 2254825 | 353366 | 147571 | 110540 | 37031 | 0.46 | 0.30 | 0.32 | 0.31 |
| 1965 | 701920 | 330960 | 140223 | 97143 | 43080 | 0.38 | 0.28 | 0.24 | 0.29 |
| 1966 | 594050 | 360172 | 166552 | 101834 | 64718 | 0.40 | 0.24 | 0.33 | 0.28 |
| 1967 | 407196 | 416311 | 163365 | 108819 | 54546 | 0.42 | 0.25 | 0.31 | 0.26 |
| 1968 | 438895 | 404080 | 139521 | 111534 | 27987 | 0.33 | 0.21 | 0.21 | 0.28 |
| 1969 | 658811 | 372570 | 142820 | 121651 | 21169 | 0.35 | 0.26 | 0.17 | 0.33 |
| 1970 | 664222 | 330537 | 159982 | 130342 | 29640 | 0.46 | 0.34 | 0.28 | 0.39 |
| 1971 | 420331 | 315802 | 136939 | 113944 | 22995 | 0.38 | 0.29 | 0.21 | 0.36 |
| 1972 | 374300 | 319302 | 142475 | 122843 | 19632 | 0.41 | 0.32 | 0.18 | 0.38 |
| 1973 | 1320356 | 269028 | 143783 | 130429 | 13354 | 0.47 | 0.41 | 0.13 | 0.48 |
| 1974 | 1136000 | 276144 | 157485 | 112540 | 44945 | 0.49 | 0.41 | 0.20 | 0.41 |
| 1975 | 864714 | 288327 | 195235 | 108536 | 86699 | 0.58 | 0.39 | 0.43 | 0.38 |
| 1976 | 691691 | 302097 | 166917 | 113670 | 53247 | 0.42 | 0.30 | 0.27 | 0.38 |
| 1977 | 990829 | 308977 | 176689 | 119188 | 57501 | 0.51 | 0.34 | 0.31 | 0.39 |
| 1978 | 920713 | 296206 | 159639 | 113984 | 45655 | 0.46 | 0.35 | 0.22 | 0.38 |
| 1979 | 905430 | 294824 | 213282 | 145347 | 67935 | 0.66 | 0.48 | 0.35 | 0.49 |
| 1980 | 1148883 | 274888 | 171844 | 140764 | 31080 | 0.55 | 0.49 | 0.15 | 0.51 |
| 1981 | 901574 | 264149 | 174264 | 141233 | 33031 | 0.54 | 0.47 | 0.16 | 0.53 |
| 1982 | 2111275 | 265691 | 205280 | 156153 | 49127 | 0.59 | 0.50 | 0.20 | 0.59 |
| 1983 | 1368338 | 325094 | 220262 | 145779 | 74483 | 0.56 | 0.45 | 0.25 | 0.45 |
| 1984 | 1299663 | 346523 | 236588 | 165772 | 70816 | 0.57 | 0.43 | 0.26 | 0.48 |
| 1985 | 1880989 | 377342 | 232387 | 171838 | 60549 | 0.51 | 0.42 | 0.21 | 0.46 |
| 1986 | 4797263 | 408832 | 308831 | 178878 | 129953 | 0.64 | 0.49 | 0.33 | 0.44 |
| 1987 | 1979144 | 481620 | 359283 | 168759 | 190524 | 0.70 | 0.50 | 0.49 | 0.35 |
| 1988 | 1830953 | 411146 | 324975 | 168552 | 156423 | 0.70 | 0.44 | 0.50 | 0.41 |
| 1989 | 1250820 | 429704 | 286684 | 178891 | 107793 | 0.61 | 0.39 | 0.44 | 0.42 |
| 1990 | 1084035 | 393180 | 240678 | 169453 | 71225 | 0.57 | 0.40 | 0.36 | 0.43 |
| 1991 | 959797 | 365941 | 238212 | 157277 | 80935 | 0.62 | 0.41 | 0.43 | 0.43 |
| 1992 | 811532 | 308734 | 193776 | 136727 | 57049 | 0.61 | 0.42 | 0.36 | 0.44 |
| 1993 | 565366 | 273504 | 163522 | 128506 | 35016 | 0.60 | 0.48 | 0.25 | 0.47 |
| 1994 | 480910 | 251627 | 145710 | 121925 | 23785 | 0.58 | 0.48 | 0.22 | 0.48 |
| 1995 | 1197928 | 236939 | 131176 | 109348 | 21828 | 0.61 | 0.54 | 0.19 | 0.46 |
| 1996 | 1339279 | 209525 | 143435 | 91386 | 52049 | 0.62 | 0.49 | 0.32 | 0.44 |
| 1997 | 2212118 | 234011 | 193103 | 92958 | 100145 | 0.77 | 0.51 | 0.64 | 0.40 |
| 1998 | 813659 | 254039 | 183561 | 79810 | 103751 | 0.68 | 0.37 | 0.55 | 0.31 |
| 1999 | 882903 | 230508 | 160702 | 89726 | 70976 | 0.67 | 0.42 | 0.36 | 0.39 |


| YEAR | RECRUITS | SSB | CATCH | LANDINGS | DISCARDS | FBAR2-6 | FBAR HC2-6 | FBAR DIS2-3 | Y/SSB |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1957 | 460518 | 274522 | 78443 | 70563 | 7880 | 0.28 | 0.23 | 0.12 | 0.26 |
| 2000 | 1035754 | 256904 | 135065 | 90754 | 44311 | 0.48 | 0.34 | 0.24 | 0.35 |
| 2001 | 577074 | 303421 | 193221 | 92912 | 100309 | 0.74 | 0.34 | 0.65 | 0.31 |
| 2002 | 1857519 | 226008 | 134277 | 79178 | 55099 | 0.56 | 0.37 | 0.38 | 0.35 |
| 2003 | 579018 | 261400 | 153997 | 74722 | 79275 | 0.60 | 0.39 | 0.41 | 0.29 |
| 2004 | 1375294 | 242136 | 127989 | 70511 | 57478 | 0.47 | 0.28 | 0.40 | 0.29 |
| 2005 | 803930 | 284394 | 119046 | 62796 | 56250 | 0.38 | 0.20 | 0.34 | 0.22 |
| 2006 | 979159 | 295586 | 131303 | 67143 | 64160 | 0.37 | 0.20 | 0.37 | 0.23 |
| 2007 | 1143879 | 300157 | 100949 | 58576 | 42373 | 0.33 | 0.17 | 0.34 | 0.20 |
| 2008 | 1071558 | 391203 | 105329 | 58336 | 46993 | 0.26 | 0.16 | 0.22 | 0.15 |
| 2009 | 1110840 | 431357 | 108262 | 62360 | 45902 | 0.23 | 0.13 | 0.21 | 0.14 |
| 2010 | 1419551 | 523991 | 116910 | 70340 | 46570 | 0.22 | 0.13 | 0.19 | 0.13 |
| 2011 | 1892434 | 507330 | 118100 | 76507 | 41593 | 0.21 | 0.12 | 0.14 | 0.15 |
| 2012 | 1274853 | 555199 | 141932 | 82018 | 59914 | 0.23 | 0.13 | 0.20 | 0.15 |
| 2013 | 1703575 | 619281 | 126247 | 86222 | 40025 | 0.19 | 0.11 | 0.15 | 0.14 |
| 2014 | 1966051 | 774978 | 133623 | 80686 | 52937 | 0.17 | 0.08 | 0.16 | 0.10 |
| 2015 | 1140208 | 754812 | 134460 | 85360 | 49100 | 0.17 | 0.08 | 0.18 | 0.11 |

Table 8.4.1. Plaice in 4 and 3.a. Input table for RCT3 analysis.

| YEARCLASS | AGE 1 XSA | AGE 2 XSA | SNSO | SNS 1 | SNS2 | BTS 1 | BTS2 | DFS0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1974 | 864714 | 548413 | NA | NA | 2402.6 | NA | NA | NA |
| 1975 | 691691 | 448206 | NA | NA | 3423.8 | NA | NA | NA |
| 1976 | 990829 | 649359 | NA | NA | 12678 | NA | NA | NA |
| 1977 | 920713 | 616240 | NA | NA | 9828.8 | NA | NA | NA |
| 1978 | 905430 | 539147 | NA | NA | 12882.3 | NA | NA | NA |
| 1979 | 1148883 | 823274 | NA | NA | 18785.3 | NA | NA | NA |
| 1980 | 901574 | 687932 | NA | NA | 8642.0 | NA | NA | NA |
| 1981 | 2111275 | 1515904 | NA | NA | 13908.6 | NA | NA | NA |
| 1982 | 1368338 | 988318 | NA | NA | 10412.8 | NA | NA | NA |
| 1983 | 1299663 | 880358 | NA | NA | 13847.8 | NA | NA | NA |
| 1984 | 1880989 | 1316261 | NA | NA | 7580.4 | NA | NA | NA |
| 1985 | 4797263 | 3276305 | NA | NA | 32991.1 | NA | NA | NA |
| 1986 | 1979144 | 1446915 | NA | NA | 14421.1 | NA | NA | NA |
| 1987 | 1830953 | 1325118 | NA | NA | 17810.2 | NA | NA | NA |
| 1988 | 1250820 | 927707 | NA | NA | 7496.0 | NA | NA | NA |
| 1989 | 1084035 | 841178 | NA | NA | 11247.2 | NA | NA | NA |
| 1990 | 959797 | 692911 | NA | NA | 13841.8 | NA | NA | 439.6 |
| 1991 | 811532 | 599095 | NA | NA | 9685.6 | NA | NA | 332.4 |
| 1992 | 565366 | 416644 | NA | NA | 4976.6 | NA | NA | 180.3 |
| 1993 | 480910 | 374746 | NA | NA | 2796.4 | NA | NA | 217.0 |
| 1994 | 1197928 | 963465 | NA | NA | 10268.2 | NA | 99.6 | 283.4 |
| 1995 | 1339279 | 1104956 | NA | NA | 4472.7 | 143.9 | 28.7 | 146.1 |
| 1996 | 2212118 | 1878375 | NA | NA | 30242.2 | 386.8 | 177.6 | 619.6 |
| 1997 | 813659 | 636602 | NA | NA | 10272.1 | 131.2 | 53.6 | 229.2 |
| 1998 | 882903 | 677250 | NA | NA | 2493.4 | 117.0 | 38.9 | NA |
| 1999 | 1035755 | 836291 | NA | 22855 | 2898.5 | 108.4 | 39.8 | NA |
| 2000 | 577074 | 488990 | 24213.5 | 11510.5 | 1102.7 | 80.3 | 26.7 | 124.9 |
| 2001 | 1857519 | 1384282 | 99628.0 | 30809.2 | NA | 217.3 | 94.4 | 313.2 |
| 2002 | 579018 | 458760 | 31202.0 | NA | 1349.7 | 53.6 | 30.3 | 122.9 |
| 2003 | 1375294 | 1021614 | NA | 18201.6 | 1818.9 | 101.4 | 45.6 | 238.6 |
| 2004 | 803930 | 635267 | 13537.2 | 10118.4 | 1571.0 | 70.8 | 42.9 | 126.7 |
| 2005 | 979159 | 675430 | 27390.6 | 12164.2 | 2133.9 | 54.9 | 44.4 | 85.9 |
| 2006 | 1143879 | 959761 | 51124.2 | 14174.5 | 2700.4 | 139.4 | 89.7 | 168.0 |
| 2007 | 1071558 | 840310 | 40580.9 | 14705.8 | 2018.7 | 98.9 | 76.5 | 98.3 |
| 2008 | 1110840 | 862227 | 50179.3 | 14860 | 1811.5 | 170.8 | 69.5 | 129.7 |
| 2009 | 1419551 | 1123228 | 53258.8 | 11946.9 | 1142.5 | 144.8 | 126.0 | 141.9 |
| 2010 | 1892434 | 1599807 | 49347.2 | 18348.6 | 2928.6 | 226.5 | 149.6 | 179.6 |
| 2011 | 1274853 | 1065903 | 52643.0 | 5893.4 | 3021.3 | 118.4 | 90.5 | 93.0 |
| 2012 | NA | NA | 45027.1 | 15394.9 | 2258.3 | 192.8 | 123.2 | 181.1 |
| 2013 | NA | NA | 44327.5 | 17312.7 | 5040.4 | 155.2 | 156.6 | 168.5 |
| 2014 | NA | NA | 11722.3 | 16726.5 | NA | 116.5 | NA | 108.0 |
| 2015 | NA | NA | 30494.5 | NA | NA | NA | NA | 100.2 |

Table 8.4.2. Plaice in 4 and 3.a. RCT3 results for age 1 in 2016 (yearclass 2015).

```
Analysis by RCT3 ver4.0
Plaice
Data for 6 surveys over 42 years : 1974 - 2015
Regression type = C
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as . 00
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.
yearclass:2015
index slope intercept se rsquare n indices pred se.pred WAP.weights
SNS0 1.006 3.230 0.368 0.5665 11 10.33 13.62 0.430 0.5039
DFS0 2.775 -0.631 1.446 0.0835 20 4.61 12.15 1.614 
VPA Mean NA NA NA NA 38 NA 13.95 0.450 0.4603
    WAP logWAP int.se
yearclass:2015 907736 13.72 0.3052
```

Table 8.4.3. Plaice in IV and 3.a. RCT3 results for age 2 in 2015 (yearclass 2013).
Analysis by RCT3 ver4.0
Plaice
Data for 6 surveys over 42 years : 1974 - 2015
Regression type = C
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as . 00
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.

```
yearclass:2014
index slope intercept se rsquare n indices pred se.pred WAP.weights
SNS0 1.012 2.939 0.371 0.5645 11 9.37 12.42 0.524 0.1488
SNS1 2.085 -6.227 0.851 0.1421 12 9.73 14.05 0.976 0.0428
```



```
DFS0 3.128 -2.716 1.641 0.0715 20 4.68 11.93 1.818 0.0124
VPA Mean NA NA NA NA 38 NA 13.65 0.457 0.1955
    WAP logWAP int.se
yearclass:2014 704951 13.47 0.202
```

Table 8.5.1. Plaice in 4 and 3.a. Input to the short term forecast ( $F$ values presented are for $F s q$ )

| $\begin{gathered} 2016 \\ \text { SSB } \end{gathered}$ | $\begin{gathered} 2016 \\ \text { F2-6 } \end{gathered}$ | $\begin{gathered} 2016 \\ \text { F_DIS2-3 } \end{gathered}$ | $\begin{gathered} 2016 \\ \text { F_HC2-6 } \end{gathered}$ | $2016$ <br> RECRUITS | $2016$ <br> LANDINGS | $2016$ <br> DISCARDS | $2016$ <br> Catch | 2016 <br> TAC |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 945709 | 0.173 | 0.16 | 0.089 | 907736 | 109277 | 42090 | 151362 | 138432 |  |  |  |
| age | year | f | f.disc | f.land | stock.n | catch.wt | landings.wt | discards.wt | stock.wt | mat | M |
| 1 | 2016 | 0.109 | 0.11 | 0 | 907736 | 0.03 | 0.08 | 0.03 | 0.04 | 0 | 0.1 |
| 2 | 2016 | 0.185 | 0.18 | 0.01 | 952339 | 0.09 | 0.25 | 0.08 | 0.09 | 0.5 | 0.1 |
| 3 | 2016 | 0.196 | 0.14 | 0.05 | 1078478 | 0.17 | 0.28 | 0.13 | 0.14 | 0.5 | 0.1 |
| 4 | 2016 | 0.192 | 0.07 | 0.13 | 826358 | 0.26 | 0.32 | 0.15 | 0.21 | 1 | 0.1 |
| 5 | 2016 | 0.16 | 0.03 | 0.13 | 423556 | 0.33 | 0.36 | 0.17 | 0.3 | 1 | 0.1 |
| 6 | 2016 | 0.135 | 0.01 | 0.13 | 515392 | 0.4 | 0.41 | 0.2 | 0.35 | 1 | 0.1 |
| 7 | 2016 | 0.094 | 0 | 0.09 | 271217 | 0.47 | 0.48 | 0.23 | 0.42 | 1 | 0.1 |
| 8 | 2016 | 0.064 | 0 | 0.06 | 151406 | 0.54 | 0.54 | 0.31 | 0.47 | 1 | 0.1 |
| 9 | 2016 | 0.041 | 0 | 0.04 | 101580 | 0.61 | 0.61 | 0.22 | 0.53 | 1 | 0.1 |
| 10 | 2016 | 0.041 | 0 | 0.04 | 206784 | 0.64 | 0.64 | 0 | 0.51 | 1 | 0.1 |
| 1 | 2017 | 0.109 | 0.11 | 0 | 980962 | 0.03 | 0.08 | 0.03 | 0.04 | 0 | 0.1 |
| 2 | 2017 | 0.185 | 0.18 | 0.01 | NA | 0.09 | 0.25 | 0.08 | 0.09 | 0.5 | 0.1 |
| 3 | 2017 | 0.196 | 0.14 | 0.05 | NA | 0.17 | 0.28 | 0.13 | 0.14 | 0.5 | 0.1 |
| 4 | 2017 | 0.192 | 0.07 | 0.13 | NA | 0.26 | 0.32 | 0.15 | 0.21 | 1 | 0.1 |
| 5 | 2017 | 0.16 | 0.03 | 0.13 | NA | 0.33 | 0.36 | 0.17 | 0.3 | 1 | 0.1 |
| 6 | 2017 | 0.135 | 0.01 | 0.13 | NA | 0.4 | 0.41 | 0.2 | 0.35 | 1 | 0.1 |
| 7 | 2017 | 0.094 | 0 | 0.09 | NA | 0.47 | 0.48 | 0.23 | 0.42 | 1 | 0.1 |
| 8 | 2017 | 0.064 | 0 | 0.06 | NA | 0.54 | 0.54 | 0.31 | 0.47 | 1 | 0.1 |
| 9 | 2017 | 0.041 | 0 | 0.04 | NA | 0.61 | 0.61 | 0.22 | 0.53 | 1 | 0.1 |
| 10 | 2017 | 0.041 | 0 | 0.04 | NA | 0.64 | 0.64 | 0 | 0.51 | 1 | 0.1 |
| 1 | 2018 | 0.109 | 0.11 | 0 | 980962 | 0.03 | 0.08 | 0.03 | 0.04 | 0 | 0.1 |
| 2 | 2018 | 0.185 | 0.18 | 0.01 | NA | 0.09 | 0.25 | 0.08 | 0.09 | 0.5 | 0.1 |
| 3 | 2018 | 0.196 | 0.14 | 0.05 | NA | 0.17 | 0.28 | 0.13 | 0.14 | 0.5 | 0.1 |
| 4 | 2018 | 0.192 | 0.07 | 0.13 | NA | 0.26 | 0.32 | 0.15 | 0.21 | 1 | 0.1 |
| 5 | 2018 | 0.16 | 0.03 | 0.13 | NA | 0.33 | 0.36 | 0.17 | 0.3 | 1 | 0.1 |
| 6 | 2018 | 0.135 | 0.01 | 0.13 | NA | 0.4 | 0.41 | 0.2 | 0.35 | 1 | 0.1 |
| 7 | 2018 | 0.094 | 0 | 0.09 | NA | 0.47 | 0.48 | 0.23 | 0.42 | 1 | 0.1 |
| 8 | 2018 | 0.064 | 0 | 0.06 | NA | 0.54 | 0.54 | 0.31 | 0.47 | 1 | 0.1 |
| 9 | 2018 | 0.041 | 0 | 0.04 | NA | 0.61 | 0.61 | 0.22 | 0.53 | 1 | 0.1 |
| 10 | 2018 | 0.041 | 0 | 0.04 | NA | 0.64 | 0.64 | 0 | 0.51 | 1 | 0.1 |

Table 8.5.2. Plaice in 4 and 3.a. Results from the short term forecast assuming $F_{2016}=F_{2015}$ (rescaled)

| BASIS | LANDINGS 2017* | F2-6 | F |  | DISCARDS 2017* | CATCH 2017 | SSB $2017$ | $\begin{gathered} S S S B \\ 2018 \end{gathered}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fmp | 184384 | 0.3 | 0.15 | 0.28 | 55227 | 239611 | 1033466 | 983389 | -5 | 28 |
| Ftar | 184384 | 0.3 | 0.15 | 0.28 | 55227 | 239611 | 1033466 | 983389 | -5 | 28 |
| Fmsy | 121523 | 0.19 | 0.1 | 0.18 | 36678 | 158201 | 1033466 | 1065323 | 3 | -15 |
| Fmsy_low | 84850 | 0.13 | 0.07 | 0.12 | 25656 | 110506 | 1033466 | 1113470 | 8 | -41 |
| Fmsy_high | 167778 | 0.27 | 0.14 | 0.25 | 50370 | 218148 | 1033466 | 1004958 | -3 | 17 |
| Fpa | 261819 | 0.45 | 0.23 | 0.41 | 77428 | 339247 | 1033466 | 883590 | -15 | 82 |
| Flim | 343667 | 0.63 | 0.32 | 0.58 | 100022 | 443689 | 1033466 | 779638 | -25 | 139 |
| SSB>Bpa | 815440 | 2.754 | 1.41 | 2.54 | 203288 | 1018728 | 1033466 | 230000 | -78 | 465 |
| SSB>Blim | 886406 | 3.57 | 1.83 | 3.29 | 212409 | 1098815 | 1033466 | 160000 | -85 | 514 |
| SSB>MSYBtrig | 815440 | 2.754 | 1.41 | 2.54 | 203288 | 1018728 | 1033466 | 230000 | -78 | 465 |
| TACsq | 143480 | 0.227 | 0.12 | 0.21 | 43207 | 186687 | 1033466 | 1036616 | 0 | 0 |
| 15\%_TAC_inc | 165142 | 0.265 | 0.14 | 0.24 | 49596 | 214738 | 1033466 | 1008386 | -2 | 15 |
| 15\%_TAC_dec | 121818 | 0.19 | 0.1 | 0.18 | 36766 | 158584 | 1033466 | 1064937 | 3 | -15 |
| Fsq* ${ }^{*}$ | 0 | 0 | NA | NA | 0 | 0 | 1033466 | 1227002 | 19 | -100 |
| Fsq* 0.25 | 28442 | 0.043 | 0.02 | 0.04 | 8437 | 36879 | 1033466 | 1187986 | 15 | -80 |
| Fsq* 0.5 | 57463 | 0.087 | 0.04 | 0.08 | 17336 | 74799 | 1033466 | 1149580 | 11 | -60 |
| Fsq* 0.9 | 100957 | 0.156 | 0.08 | 0.14 | 30514 | 131471 | 1033466 | 1092293 | 6 | -29 |
| Fsq* ${ }^{*}$ | 111309 | 0.173 | 0.09 | 0.16 | 33623 | 144932 | 1033466 | 1078707 | 4 | -22 |
| Fsq* 1.1 | 122119 | 0.191 | 0.1 | 0.18 | 36856 | 158975 | 1033466 | 1064542 | 3 | -15 |
| Fsq* ${ }^{*} .25$ | 137462 | 0.217 | 0.11 | 0.2 | 41423 | 178885 | 1033466 | 1044474 | 1 | -4 |

Wanted catch of plaice in Subarea 4 and Subdivision 3.a.20, calculated as the projected total stock wanted catch less the wanted catch of plaice from Subarea 4 taken in Division 7.d. The subtracted value ( 934 t ) is estimated based on the plaice catch advice for Division 7.d for 2016, using the recent 10-year average (2006-2015) proportion of plaice from Subarea 4 in the annual plaice landings in Division $7 . d$. Similarly, 652 t of unwanted catch of plaice from Subarea 4 are projected to be taken in Division 7.d. These are removed from the unwanted catch. TAC change restrictions of $15 \%$ are applied after subtracting the Division 7.d catches.
**SSB 2018 relative to SSB 2017
***landings2017/(131714+11766), where 131714 and 11766 are the TAC for 2016 Subarea 4 and Division 7.d.

Table 8.5.3. Plaice in 4 and 3.a. Detailed STF table by age, assuming F = Fsq, rescaled.

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | F | DISC | Lan | .N | .wT | .wT | .wT | .wT | .N | CATCH | .N | LAND | .N | DIsc | SSB | TSB |
| 2016 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.11 | 0.11 | 0 | 907736 | 0.03 | 0.08 | 0.03 | 0.04 | 88934 | 3021 | 127 | 10 | 88806 | 3019 | 0 | 34797 |
| 2 | 0.19 | 0.18 | 0.01 | 952339 | 0.09 | 0.25 | 0.08 | 0.09 | 153345 | 13537 | 5277 | 1311 | 148067 | 12240 | 43808 | 87615 |
| 3 | 0.20 | 0.14 | 0.05 | 1078478 | 0.17 | 0.28 | 0.13 | 0.14 | 182933 | 31225 | 51221 | 14513 | 131713 | 16727 | 77471 | 154941 |
| 4 | 0.19 | 0.07 | 0.13 | 826358 | 0.26 | 0.32 | 0.15 | 0.21 | 137307 | 35916 | 90379 | 28680 | 46929 | 7227 | 169954 | 169954 |
| 5 | 0.16 | 0.03 | 0.13 | 423556 | 0.33 | 0.36 | 0.17 | 0.30 | 59670 | 19649 | 48953 | 17803 | 10716 | 1836 | 128620 | 128620 |
| 6 | 0.14 | 0.01 | 0.13 | 515392 | 0.4 | 0.41 | 0.20 | 0.35 | 61902 | 24745 | 57600 | 23885 | 4301 | 847 | 181590 | 181590 |
| 7 | 0.09 | 0 | 0.09 | 271217 | 0.47 | 0.48 | 0.23 | 0.42 | 23171 | 11000 | 22486 | 10838 | 686 | 160 | 113188 | 113188 |
| 8 | 0.06 | 0 | 0.06 | 151406 | 0.54 | 0.54 | 0.31 | 0.47 | 8939 | 4831 | 8836 | 4801 | 103 | 32 | 71262 | 71262 |
| 9 | 0.04 | 0 | 0.04 | 101580 | 0.61 | 0.61 | 0.22 | 0.53 | 3883 | 2352 | 3883 | 2352 | 0 | 0 | 53668 | 53668 |
| 10 | 0.04 | 0 | 0.04 | 206784 | 0.64 | 0.64 | 0 | 0.51 | 7904 | 5085 | 7904 | 5085 | 0 | 0 | 106149 | 106149 |
| 2017 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.11 | 0.11 | 0 | 980962 | 0.03 | 0.08 | 0.03 | 0.04 | 96108 | 3265 | 138 | 11 | 95970 | 3263 | 0 | 37604 |
| 2 | 0.19 | 0.18 | 0.01 | 736869 | 0.09 | 0.25 | 0.08 | 0.09 | 118650 | 10474 | 4083 | 1014 | 114567 | 9471 | 33896 | 67792 |
| 3 | 0.20 | 0.14 | 0.05 | 716131 | 0.17 | 0.28 | 0.13 | 0.14 | 121471 | 20734 | 34012 | 9637 | 87460 | 11107 | 51442 | 102884 |
| 4 | 0.19 | 0.07 | 0.13 | 802191 | 0.26 | 0.32 | 0.15 | 0.21 | 133292 | 34865 | 87735 | 27841 | 45556 | 7016 | 164984 | 164984 |
| 5 | 0.16 | 0.03 | 0.13 | 617372 | 0.33 | 0.36 | 0.17 | 0.3 | 86974 | 28641 | 71354 | 25949 | 15620 | 2676 | 187475 | 187475 |
| 6 | 0.14 | 0.01 | 0.13 | 326589 | 0.40 | 0.41 | 0.20 | 0.35 | 39225 | 15680 | 36500 | 15135 | 2726 | 537 | 115068 | 115068 |
| 7 | 0.09 | 0 | 0.09 | 407554 | 0.47 | 0.48 | 0.23 | 0.42 | 34819 | 16529 | 33789 | 16286 | 1030 | 241 | 170086 | 170086 |
| 8 | 0.06 | 0 | 0.06 | 223392 | 0.54 | 0.54 | 0.31 | 0.47 | 13189 | 7129 | 13037 | 7083 | 152 | 47 | 105143 | 105143 |
| 9 | 0.04 | 0 | 0.04 | 128503 | 0.61 | 0.61 | 0.22 | 0.53 | 4912 | 2975 | 4912 | 2975 | 0 | 0 | 67893 | 67893 |
| 10 | 0.04 | 0 | 0.04 | 267815 | 0.64 | 0.64 | 0 | 0.51 | 10237 | 6586 | 10237 | 6586 | 0 | 0 | 137479 | 137479 |
| 2018 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.11 | 0.11 | 0 | 980962 | 0.03 | 0.08 | 0.03 | 0.04 | 96108 | 3265 | 138 | 11 | 95970 | 3263 | 0 | 37604 |
| 2 | 0.19 | 0.18 | 0.01 | 796312 | 0.09 | 0.25 | 0.08 | 0.09 | 128222 | 11319 | 4413 | 1096 | 123809 | 10235 | 36630 | 73261 |
| 3 | 0.20 | 0.14 | 0.05 | 554104 | 0.17 | 0.28 | 0.13 | 0.14 | 93988 | 16043 | 26316 | 7456 | 67672 | 8594 | 39803 | 79606 |
| 4 | 0.19 | 0.07 | 0.13 | 532671 | 0.26 | 0.32 | 0.15 | 0.21 | 88508 | 23151 | 58258 | 18487 | 30250 | 4659 | 109553 | 109553 |
| 5 | 0.16 | 0.03 | 0.13 | 599317 | 0.33 | 0.36 | 0.17 | 0.30 | 84430 | 27803 | 69267 | 25190 | 15163 | 2598 | 181992 | 181992 |
| 6 | 0.14 | 0.01 | 0.13 | 476033 | 0.40 | 0.41 | 0.20 | 0.35 | 57174 | 22856 | 53202 | 22061 | 3973 | 783 | 167722 | 167722 |
| 7 | 0.09 | 0 | 0.09 | 258255 | 0.47 | 0.48 | 0.23 | 0.42 | 22064 | 10474 | 21411 | 10320 | 653 | 153 | 107778 | 107778 |
| 8 | 0.06 | 0 | 0.06 | 335689 | 0.54 | 0.54 | 0.31 | 0.47 | 19819 | 10712 | 19590 | 10644 | 229 | 71 | 157997 | 157997 |
| 9 | 0.04 | 0 | 0.04 | 189600 | 0.61 | 0.61 | 0.22 | 0.53 | 7248 | 4390 | 7248 | 4390 | 0 | 0 | 100172 | 100172 |
| 10 | 0.04 | 0 | 0.04 | 344204 | 0.64 | 0.64 | 0 | 0.51 | 13157 | 8465 | 13157 | 8465 | 0 | 0 | 176691 | 176691 |

## Catche, Landing and Discards



Figure 8.2.1 North Sea plaice (including Skagerrak and 7.d Q1). Time series of catch (dashed line), landings (solid line) and discards (gray line) estimates.

Discards at age


Landings at age


Figure 8.2.2 North Sea and Skagerrak plaice (including 7.dQ1). Discards numbers-at-age (top) and landing numbers-at-age (down).

## Catches at age



Figure 8.2.3 North Sea and Skagerrak plaice (including 7.dQ1). Catch numbers-at-age.


Figure 8.2.4 North Sea and Skagerrak plaice. Stock weight-at-age (top left), landings weight-at-age (top right), discards weight-at-age (bottom left) and catch weight-at-age (bottom right).


Figure 8.2.5 North Sea and Skagerrak plaice. Standardized survey tuning indices used for tuning XSA: BTS-combined (red), BTS-Isis-old (black) SNS-1 (1984-1999, blue) and SNS-2 (2000-2015, grey). Note: only ages used in the assessment are presented. The BTS-combined index combines BTS-Tridens and BTS-Isis indices.

## BTS-Combined



Figure 8.2.6 Plaice in 4 and 3.a. Internal consistency plot for the BTS-combined survey index.


Figure 8.2.7. Plaice in 4 and 3.a. Internal consistency plot for the BTS-Isis-early survey index.


Figure 8.2.8. Plaice in 4 and 3.a. Internal consistency plot for the SNS-1 (1984-1999, left) and the SNS-2 (2000-2015, right) survey indices.

(a)

(b)

Figure 8.2.9. Data upload in Intercatch: percentage of landings (a) and discards (b) (\%) by country by métier. Sampled and unsampled refers to availability of age-composition information. This data is for North Sea plaice only (excluding Skagerrak).


Figure 8.3.1 North Sea and Skagerrak plaice. Standardized survey tuning indices for the surveys used in the recent part of the assessment and the two vessels participating in the industry survey.


Figure 8.3.2 North Sea and Skagerrak plaice. SSB(top left), fishing mortality(top right), and recruitment (bottom left) estimates of the assessment with the inclusion of the industry survey (dashed line), compared with the assessment with the indices as used in WGNSSK2015 (drawn line.


Figure 8.3.3. Plaice in 4 and 3.a. Log catchability residuals for the final XSA run from the three tuning series.


Figure 8.3.4. North Sea and Skagerrak plaice. Retrospective pattern of the final XSA run with respect to SSB, recruitment and F.


Figure 8.4.1. North Sea and Skagerrak plaice. Stock summary figure, time series on SSB (top left), fishing mortality for ages 2-6 (top right), and recruitment (bottom left).


Figure 8.6.1. North Sea and Skagerrak plaice. Stock summary figure, time series on SSB (top left), fishing mortality for ages 2-6 (top right), and recruitment (bottom left).

## $9 \quad$ Sole (Solea solea) in Division 7.d (Eastern English Channel)

The assessment of sole in sub-area 7.d is presented here as an update assessment.
All the relevant biological and methodological information can be found in the Stock Annex dealing with this stock. Here, only the basic input and output from the assessment model will be presented.

### 9.1 General

### 9.1.1 Biology and ecosystem aspects

No new information on ecosystem aspects was presented at the working group in 2016.
All available information on ecological aspects can be found in the Stock Annex.

### 9.1.2 Stock identity and possible assessment areas

Reflections on the stock identity of Eastern Channel sole can be consulted in the Stock Annex.

More recently, Cuveliers et al. (2012) showed genetic differences (using neutral microsatellite markers and a mitochondrial marker) at a large scale, along a latitudinal gradient from the Skagerrak/Kattegat to the Bay of Biscay. At a smaller spatial scale within the North Sea Ecoregion however, subpopulations seemed genetically homogeneous, probably due to a high level of gene flow and/or the high effective population size preventing strong effects of genetic drift. With respect to the temporal aspect, a remarkable high genetic stability was found from the 1950s up to present (Cuveliers et al. 2011). The large scale genetic differentiation supports the approach in which sole is treated as several different assessment and management units within the North Sea Ecoregion, whereas the homogeneity suggests it may be one population.

### 9.1.3 Management regulations

Management of sole in 7.d is by TAC and technical measures. The minimum landing size for sole is 24 cm . Mesh size restrictions in place are 80 mm for beam trawling and 90 mm for otter trawlers. Fixed nets are required to use 100 mm mesh since 2002 although an exemption to permit 90 mm has been in force since that time.

TACs have been defined for sole in 7.d. A historical overview since 2000 is presented in the table below.

Historical overview of the TACs for sole Solea solea in Division 7.d, 2000-2015

| YeAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| TAC | 4100 | 4600 | 5200 | 5400 | 5900 | 5700 | 5720 | 6220 | 6590 |
| Year | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |  |
| TAC | 5274 | 4219 | 4852 | 5580 | 5900 | 4838 | 3483 | $3258^{*}$ |  |

* Catch TAC (see below)

Except for 2010, the TAC was not restrictive for France, Belgium or the UK since 1997. In 2014 it became restrictive for Belgium, and in 2015 this was the case for Belgium and France (see 9.2.1 Landings).

In the second half of 2015, the North Western Waters Advisory Council (in this case mainly driven by France and Belgium) approached the Commission with a management strategy in which it was found to be appropriate (STECF evaluation) to set a TAC of 3000 tonnes for 2016 (the ICES advice was for landings of 2376 tonnes in 2016), corresponding to a $14 \%$ decrease as compared to 2015. The Commission and the Member States concerned agreed that the following additional rules should be considered in future years unless scientific advice indicates that they are no longer appropriate: i) keep the TAC constant at 3000 tonnes in 2016-2020, ii) if the biomass in any year before 2020 is below the precautionary level (Bpa), then the TAC will be set at a level corresponding to a fishing mortality equal to FMSY and iii) if ICES indicates in 2019 that the fishing mortality in 2020 risks being above $\mathrm{F}_{\text {mSY, }}$ then the TAC will be set at a level corresponding to a fishing mortality in line with $\mathrm{F}_{\mathrm{msy}}$. If the fishing mortality is below Fmsy for any 2 consecutive years before 2020 then the Commission will request the STECF to provide advice on the situation of this stock.

As 7.d sole is under the landing obligation in 2016, the landings TAC of 3000 tonnes was topped up to a catch TAC of 3258 tonnes based on a discard ratio of $8,6 \%$. It is however unclear to the stock coordinator and the WGNSSK 2016 how this discard ratio was derived, as it differs from the values estimated by the WG in 2015. So far, this management strategy is not anchored in EU legislation, and was not brought to the attention of ICES. Therefore, it cannot be used as the basis for the ICES advice. The corresponding catch options are not included in the advice sheet.
In response to the drop in SSB and the poor recruitment in 2012-2014, the two main countries participating in the fishery have also implemented additional conservation measures. For Belgian beam trawlers in $7 . \mathrm{d}$ (and $7 \mathrm{fg}, 7 \mathrm{a}$ ) it is mandatory since 1 April 2015 to incorporate a 3 m long section with 120 mm mesh size before the codend, in order to reduce the catches of small sole. France engaged in 2016 to i) strengthen the protection of the nursery areas, ii) increase the area closed to fishing within the nursery areas, and iii) increase the minimum conservation reference size to 25 cm for French vessels in accordance with EU legislation, where appropriate.

### 9.1.4 ICES advice

In $\underline{2014}$ the stock status was presented as follows:


The ICES advice for 2015 was:
ICES advises on the basis of the MSY approach but cannot quantify the resulting catches. The implied landings should be no more than 1931 t .

In $\underline{2015}$ the stock status was presented as follows:

|  | Fishing pressure |  |  |  |  | Stock size |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 2012 | 2013 |  | 2014 |  | 2013 | 2014 |  | 2015 |
| Maximum Sustainable Yield | $\mathrm{F}_{\mathrm{MSY}}$ | $\bigcirc$ | * |  | Above | MSY <br> $\mathrm{B}_{\text {trigeer }}$ |  |  |  | At trigger |
| Precautionary approach | $\begin{aligned} & \mathrm{F}_{\mathrm{paa}}, \\ & \mathrm{~F}_{\mathrm{llm}} \end{aligned}$ |  | ) |  | Harvested unsustainably | $\mathrm{B}_{\mathrm{pa}}, \mathrm{Blim}$ |  | ( |  | Full reproductive capacity |
| Management Plan | $F_{\text {mGt }}$ | - | - |  | Not applicable | SSB mat $^{\text {m }}$ | - | - | - | Not applicable |

The ICES advice for 2016 was:
ICES advises that when the MSY approach is applied, catches in 2016 should be no more than 2685 tonnes. If this stock is not under the EU landing obligation in 2016 and discard rates do not change from 2014, this implies landings of no more than 2376 tonnes.

### 9.2 Data

A detailed description of the fishery can be found in the Stock Annex.
There were no revisions to the landings data provided last year.

### 9.2.1 Landings

Table 9.1 and Figure 9.1 summarise the official sole landings by country for Division 7.d, as consulted in ICES Fishstat, and the total ICES estimated landings (for 2012-2015, these are the landings uploaded to InterCatch by the countries involved in the fishery). The landings have steadily increased over the ' 70 s - ‘ 90 s, fluctuated around an average of 4815 t (range: $3832 \mathrm{t}-6247 \mathrm{t}$ ) in 2000-2014, and dropped to 3372 tonnes in 2015 . Over the last ca 30 years, the contribution to the landings of the three main countries involved in this fishery has remained rather stable over time (around 30\% Belgium, 20\% UK, and 50\% France) (Figure 9.2).

Since 1997, full uptake of the 7.d sole TAC has not been realized, and also the national quota have not been restrictive. In 2014 however, the national Belgian quotum was overshot by $15 \%$. In 2015 both Belgium and France overshot their national quota (Belgium by $12 \%$, France by $3 \%$ ). The total uptake was $99 \%$ in this year (for comparison: $73 \%$ in $2012,74 \%$ in $2013,95 \%$ in 2014). The 2015 uptake percentages (both national and total) are presented in Table 9.2 and Figure 9.3.
The increase in Belgian effort (mainly due to Belgian beam trawlers trying to 'escape' the sharp increase in Dutch pulse trawler effort in the southern North Sea) also becomes apparent when comparing the landings by gear type between 2013 and 2014, that show an increase from $26 \%$ to $35 \%$ for TBB, the gear mainly used by Belgian fishermen. In 2015, the relative importance of the different métiers landing sole from 7.d remained almost identical to 2014 (Table 9.3).

The 2016 ICES Data Call detailed that biological data accompanying the landings should be uploaded to InterCatch for the year 2015. In this way, biological data (age and length distributions) were available for $87 \%$ of the landings. Figures 9.4 and 9.5 illustrate the distribution of these data by métier/country and country respectively.

### 9.2.2 Discards

Until the 2014 meeting of the WGNSSK, it was decided not to include discards in the assessment of 7.d sole due to the scarcity of the data. Furthermore, with an estimated overall discard rate of around $10 \%$ (2011-2013), discards were considered not to be a substantial part of the catch for this high valued species (ICES 2014a).
In 2015, quarterly discard data from the different countries contributing to the sole fisheries in 7.d were requested through InterCatch for the years 2012-2014, and gaps were filled by allocating discard percentages from similar métiers/quarters to the unsampled strata. Only discard data obtained through InterCatch have been used in the WGNSSK 2015, making 2015 the first year in which InterCatch was effectively used for the collection and raising of $7 . d$ sole discards (details on older available discard information for this stock can be consulted in ICES 2014a). In the assessment year 2016, discard data for 2015 were added in the same way. $42 \%$ of the discards were observed discards, $58 \%$ was raised (Table 9.4 - this table indicates $43 \%$ observed discards ( $3 \%$ with sampled distribution, $40 \%$ with estimated distribution) but this is a rounding effect).

Discard rates (based on weight data) that were evaluated as reliable for the three countries (Belgium, France and UK) in 2012-2015 are summarized in Table 9.5. From this table it becomes clear that the different métiers contributing to this fishery were not well covered for the years 2012 and 2013 (mainly for France and UK), so only the 2014 discard rate (all main métiers sampled by all countries) was used for topping up the landings advice to catch advice in 2015. Because of sufficient coverage in 2014 and 2015, the average discard rate 2014-2015 was used for the same purpose during WGNSSK 2016 (overall discard rate of 9,25\%).

### 9.2.3 Catch at age / Weigth at age

Catch proportions at age and standardized catch proportions at age are depicted in Figures 9.6 and 9.7 respectively.
Weight at age in the catch is presented in Table 9.6 and Figure 9.8 and weight at age in the stock in Table 9.7 and Figure 9.9. The procedure for calculating mean weights is described in the Stock Annex.

### 9.2.4 Maturity and natural mortality

As in previous assessments, a knife-edged maturity-ogive was used at age 3.
Natural mortality is assumed to be a fixed value (0.1) for all ages across all years.

### 9.2.5 Tuning series

Two commercial (both beam trawl: Belgian CBT and UK(E\&W)-CBT) and three survey (UK(E\&W)-BTS-Q3, UK(E\&W)-YFS, FR-YFS) data series are used for the calibration of the assessment of 7.d sole. The full series are presented in Tables 9.8-12.

The UK survey component of the Young fish survey (YFS) was last conducted in 2006. In the absence of any update of the UK component, it was decided at the Benchmark working group WKFLAT (ICES 2009) that the UK component should still be used in the assessment independently from the French component of the YFS index. It was also noted that the lack of information from the UK YFS will affect the quality of the recruitment estimates and therefore the forecast.

Two revisions of previously submitted data were received in 2016:

## UK(E\&W)-BTS-Q3

This revision affected the entire time series, and was performed after CEFAS carried out investigations in relation to the survey data quality, selection of prime stations, age-at-length keys and ultimately the indices calculations. More details, including results for revisited index and previous index are shown in Silva (2016) in Annex 08 to facilitate a better comparison and further inform on temporal differences to the year class abundance.

## Belgian CBT

This revision affected the years 2012-2014, and was necessary as it was discovered that the tuning data for these years were not based on the exact same set of vessels as was the case previously. This was corrected, and the 2015 data were calculated in the same way.

To gain insight in the impact of these revisions on the assessment results, separate assessment runs were carried out using the same data and settings as in WGNSSK 2015 (so only data up to data year 2014, only differing from the original input data in the tuning series).

Run 1 : revision of UK(E\&W)-Q3-BTS - corresponding residual patterns, assessment summary graphs and retrospective patterns are shown in Figures 9.10-9.12, and the corresponding assessment summary numbers in Table 9.13 (WGNSSK 2015 = OLD ; WGNSSK 2015 + update = NEW).

Results: impact generally only detectable in recent years, retros similar recruitment 2014 was revised downwards

F 2014 was revised upwards
SSB 2014 was revised downwards
Run 2 : revision of BEL-CBT - corresponding residual patterns, assessment summary graphs and retrospective patterns are shown in Figures 9.13-9.15, and the corresponding assessment summary numbers in Table 9.14 (WGNSSK 2015 = OLD ; WGNSSK 2015 + update $=$ NEW).

Results: impact generally only detectable in recent years, retros similar
recruitment 2014 was revised downwards
F 2014 was revised downwards
SSB 2014 was revised upwards
Run 3 : both revisions - corresponding residual patterns, assessment summary graphs and retrospective patterns are shown in Figures 9.16-9.18, and the corresponding assessment summary numbers in Table 9.15 (WGNSSK 2015 = OLD ; WGNSSK 2015 + update $=$ NEW).

Results: impact generally only detectable in recent years, retros similar
recruitment 2014 was revised downwards (25732 -> 20574)
F 2014 was revised downwards ( $0.55->0.48$ )
SSB 2014 was revised upwards (9052 t -> 9968 t)

### 9.3 Analyses of stock trends

### 9.3.1 Review of last year's assessment

No major deficiencies for the sole assessment in the Eastern English Channel were reported.

### 9.3.2 Exploratory catch at age analysis

Catch at age analysis was carried out according to the specifications in the Stock Annex. The model used was XSA.

The time series of the standardized indices for ages 1 to 10 from the five tuning fleets (BE-CBT, UK(E\&W)-CBT, UK(E\&W)-BTS, UK(E\&W)-YFS and the FR-YFS) are plotted in Figure 9.19. All tuning fleets appear to track the year classes reasonably well for ages 2 to 6 .

Internal consistency plots for the 2 commercial fleets and the UK beam trawl survey are presented in Figures 9.20-22. The internal consistency of these three fleets is high for the entire age-range.
The catchability residuals for the proposed final XSA (see below) are shown in Figure 9.23. Some concern rises around the UK(E\&W)-BTS-Q3, that shows an age effect for Age 1 (that is more effectively estimated by the UK(E\&W)-YFS and the FR-YFS) and a year effect in 2014 (note that these are persistent effects, that were not affected by the revision of this tuning series).

In this year's assessment the estimates for the recruiting year class 2014 were estimated by the UK beam trawl survey and the French component of the Young Fish Survey which have weightings of $44,2 \%$ and $44,1 \%$ respectively in the final survivor estimates (Table 9.16). Shrinkage takes $11,7 \%$ of the weighting. Although it should be noted that the internal standard errors of both surveys are around 1.0, indicating a high variability and therefore an uncertain estimate providing for this year class strength, the Expert group decided to use this estimate in the forecast.

At age 2, the 2013 year-class is predominantly estimated by the UK beam trawl survey and the commercial UK beam trawl fleet, with weightings of $41,5 \%$ and $42,4 \%$ respectively (Table 9.16). Especially the UK commercial tuning fleet estimates the survivors of that year class to be relatively weak 8277). The Belgian commercial beam trawl fleet estimate this year class to be even weaker (3719) with a weighting of $13,1 \%$.
$F$ shrinkage gets low weights for all ages older than 2 . The weighting of the 3 surveys decreases for the older ages as the commercial fleets are given more weight.

## Deviation from the stock annex : change of F-range

WGNSSK 2015 decided to change the F bar-range from $^{2-8}$ to 3-7 in the final XSA, and as such deviated from the stock annex. The reason was that F increased to a very high value of almost 0,8 in 2014 (steep increase compared to the already high $F$ in the previous years), and this could not be explained by any anomalies in the analyses, but was identified to be due to the yearclass 2006 (8 years old in 2014) that had almost disappeared from the stock.
This approach was followed again in WGNSSK 2016. It was decided that the discussion to go back to the old age range, and around whether the reference points should be reestimated under this new setup (or whether stronger shrinkage would help out), are questions to be addressed during the benchmark of this stock in 2017.

### 9.3.3 Final assessment

The final settings used in this year's assessment are specified as in the stock annex and are detailed below:

|  | 2016 ASSESSMENT |  |  |
| :--- | :--- | :--- | :--- |
| Fleets | Years | Ages | $\alpha-\beta$ |
| BE-CBT commercial | $86-15$ | $2-10$ | $0-1$ |
| UK(E\&W)-CBT commercial | $86-15$ | $2-10$ | $0-1$ |
| UK(E\&W)-BTS survey | $88-15$ | $1-6$ | $0.5-0.75$ |
| YFS - survey (combined index UK-FR) |  |  |  |
| UK-YFS - survey | $87-06$ | $1-1$ | $0.5-0.75$ |
| FR-YFS - survey | $87-15$ | $1-1$ | $0.5-0.75$ |
|  |  |  |  |
| -First data year | 1982 |  |  |
| -Last data year | 2015 |  |  |
| -First age | 1 |  |  |
| -Last age | $11+$ |  |  |
| Time series weights | None |  |  |
| -Model | No Power model |  |  |
| -Q plateau set at age | 7 |  |  |
| -Survivors estimates shrunk towards mean F | 5 years / 5 ages |  |  |
| -s.e. of the means | 2.0 |  |  |
| -Min s.e. for pop. Estimates | 0.3 |  |  |
| -Prior weighting | None |  |  |

The diagnostics of this run (including fishing mortalities and stock numbers by age and year) are presented in Table 9.16. A summary of the XSA results is given in Table 9.17 and trends in yield, fishing mortality, recruitment and spawning stock biomass are shown in Figure 9.26.

Retrospective patterns for the final run are shown in Figure 9.27. There is a small retrospective pattern in F (that was overestimated in 2012-2013), but the consistency between estimates in successive years for fishing mortality and SSB was found to be generally good.

### 9.3.4 Historical stock trends

Trends in landings, $\mathrm{SSB}, \mathrm{F}(3-8)$ and recruitment are presented in Table 9.17 and Figure 9.26 .

For most of the time series, fishing mortality has been fluctuating between $\mathrm{F}_{\mathrm{pa}}(0.4)$ and $\mathrm{F}_{\text {lim }}$ (0.55). In the early 90 's it dropped below $\mathrm{F}_{\text {pa }}$. Since 1999 it decreased steadily from 0.55 to around 0.4 in 2001 after which it remained stable until 2005. In the last 9 years fishing mortality has fluctuated again but consistently remained above $\mathrm{F}_{\mathrm{pa}}$ (with an especially strong increase in 2013-2014).
Recruitment has fluctuated around 24 million recruits with occasional strong year classes. Five of the highest values in the time series have been recorded in the last 12 years. The 2011 and 2012 year classes were predicted to be very weak.

The spawning stock biomass has been stable for most of the time series. Since 2001 SSB has increased to well above $B_{p a}(8000 \mathrm{t})$ due to average and above average year classes. The incoming very weak year classes of 2011 and 2012 have reversed the increasing trend in SSB, that is predicted to drop below MSY B trigger in 2016.

### 9.4 Recruitment estimates and short-term forecast

### 9.4.1 Recruitment estimates

The table below summarizes the recruitment estimates for the year classes 2014-2016 from the XSA, an RCT3-analysis and the long-term geometric mean recruitments. The stock annex prescribes that the XSA value should be taken for the initial stock size for age 2 (in 2015) and older, and from RCT3 for age 1, if appropriate. Otherwise the XSA value for age 1 is used. The long-term geometric mean recruitment is used for age 0 in all projection years.

The RCT3 input for Ages 1 and 2 is presented in Tables 9.18 and 9.19 respectively, and the corresponding outputs in Tables 9.20 and 9.21, but the results have not been used for prediction.

After discussion, the WGNSSK decided to follow the stock annex and use the XSA result for yearclass 2014, the fallback option of the XSA result for yearclass 2015 (and not the RCT3), and the GM for yearclass 2016.

Since 2004 initial stock size for age 2 was taken from XSA.

| Year class | AGE in 2015 | XSA SURVIVors | RCT3 | GM1982-2012 | ACCEPTED ESTIMATE |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 2014 | 2 | 23301 | 25234 | 21074 | XSA |
| 2015 | 1 | NA | 22439 | 23722 | XSA |
| 2016 | 0 | NA | NA | 23722 | GM |

### 9.4.2 Short-term forecast

The short term prognosis was carried out according to the specifications in the stock annex. As fishing mortality has fluctuated in the last three years, the selection pattern for prediction has been taken as a 3 year unscaled average. Weights at age in the catch and in the stock are averages for the years 2013-2015.

2016 was the second year in which the short-term forecast was performed using an Rscript that uses the output files from the XSA (and the RCT3-analysis, when used) as input files, so no separate input files for the short-term forecast are presented in this report.

As the TAC was nearly fully taken in 2015 (uptake of 99\%), the TAC-constraint scenario was used for the forecast. The results for different management options under this scenario are presented in Table 9.22, and the accompanying relative contributions of yearclasses to the landings in 2017 and to SSB in 2018 are shown in Figures 9.28 and 9.29.

## ICES advice 2016

Assuming a TAC-constraint scenario and an overall discard rate of 9,25\%, ICES advises that when the MSY approach is applied, catches in 2017 should be no more than 2487 tonnes.

### 9.5 Medium- and long-term forecasts

This year, no medium- and long-term forecasts have been carried out for this stock.

### 9.6 Biological reference points

The table below summarizes all known reference points for sole in 7.d, and their technical basis.

| Framework | Reference point | Value | Technical basis |
| :---: | :---: | :---: | :---: |
| MSY <br> approach | MSY <br> BTrigger | 8000 t | Bpa |
|  | FMSY | 0.3 | Stochastic simulations assuming a smooth hockeystick relationship. Calculated on ages 3-8. |
| Precautionary approach | Blim | Not defined. | Poor biological basis for definition. |
|  | Bpa | 8000 t | This is the lowest observed biomass at which there is no indication of impaired recruitment. Smoothed Bloss. |
|  | Flim | 0.55 | Floss, but poorly defined; analogy to North Sea and setting of $1.4 \mathrm{Fpa}=0.55$. This is a fishing mortality at or above which the stock has shown continued decline. |
|  | Fpa | 0.4 | Between Fmed and 5th percentile of Floss; SSB $>$ Bpa and probability (SSBmt<Bpa), 10\%: 0.4. |
| Management plan | SSBMGT | Undefined |  |
|  | FMGT | Undefined |  |

Note that the MSY reference points have been redefined, and that $\mathrm{F}_{\text {msy }}$ was changed for this stock from 0.29 to 0.3 (ICES 2014b). MSYB ${ }_{\text {trigger }}$ remained unchanged.

### 9.7 Quality of the assessment

The main quality issue related to the evaluation of the stock status of sole in 7.d is that the UK component of the YFS index is not available since 2007, resulting in the unavailability of the combined YFS-index. This combined index estimated the incoming year class strength very consistently and provided reliable estimates for the forecasts. Although results of using the YFS indices separately (FR-YFS for 1987-present and UKYFS for 1987-2006), did not show apparent changes in retrospective patterns, it was noted that the lack of information from the UK YFS will affect the quality of the recruitment estimates and therefore the forecast. The Working Group suggests that the assessment could benefit if the French Young Fish survey could be extended to include some of the sampling points from the former UK Young Fish survey along the English coast. The extended French survey might then mimic the earlier available combined Young Fish survey which was an excellent estimator of the incoming recruitment.

Other quality-issues have been thoroughly listed in the report of WGNSSK 2014 (ICES 2014).

### 9.8 Management considerations

- There is misreporting from adjacent areas. The Working group has addressed this by modifying landings data accordingly. Since 2002 the

UK (E\&W) beam trawl landings from two rectangles 28E8 and 29E8 (in 7.d) were re-allocated to 7 e on a quarterly basis, (based on information provided to the Working Group by the fishing industry) and the age compositions raised accordingly.

- There is a less than $5 \%$ probability that SSB will decrease to $B_{p a}$ in the short term due to the strong 2008 and 2009 year classes.
- EU Council Regulation (EC) $\mathrm{N}^{\circ} 43 / 2014$ allocates different amounts of $\mathrm{Kw}^{*}$ days by Member State and area to different effort groups of vessels depending on gear and mesh size. This regime has only slightly reduced effort directed at sole in this area.
- Due to the minimum mesh size $(80 \mathrm{~mm})$ in the mixed beam trawl fishery, a large number of (undersized) plaice are discarded. The $80-\mathrm{mm}$ mesh size is matched to the minimum landing size of sole but not matched to the minimum landing size of plaice. Measures to reduce discarding of plaice in the sole fishery would greatly benefit the plaice stock and future yields. Mesh enlargement would reduce the catch of undersized plaice, but would also result in loss of marketable sole. An increase in the minimum landing size of sole could provide an incentive to fish with larger mesh sizes and therefore mean a reduction in the discarding of plaice.


### 9.9 References

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Silva, J.F. 2016. Plaice (Pleuronectes platessa) and sole (Solea solea) in the UK Beam trawl survey in the Eastern English Channel (7.d). Working document to the WGNSSK, April 2016.

Table 9.1 Sole 7.d. Nominal landings (tonnes) by country over the period 1974-2015, as officially reported to ICES and used by the Working Group.

|  |  |  |  |  |  |  | Total used |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Belgium | France | UK(E+W) | other | reported | Unallocated* | by WG |
| 1974 | 159 | 383 | 309 | 3 | 854 | 30 | 884 |
| 1975 | 132 | 464 | 244 | 1 | 841 | 41 | 882 |
| 1976 | 203 | 599 | 404 | . | 1206 | 99 | 1305 |
| 1977 | 225 | 737 | 315 | . | 1277 | 58 | 1335 |
| 1978 | 241 | 782 | 366 | . | 1389 | 200 | 1589 |
| 1979 | 311 | 1129 | 402 | . | 1842 | 373 | 2215 |
| 1980 | 302 | 1075 | 159 | . | 1536 | 387 | 1923 |
| 1981 | 464 | 1513 | 160 | . | 2137 | 340 | 2477 |
| 1982 | 525 | 1828 | 317 | 4 | 2674 | 516 | 3190 |
| 1983 | 502 | 1120 | 419 | . | 2041 | 1417 | 3458 |
| 1984 | 592 | 1309 | 505 | . | 2406 | 1169 | 3575 |
| 1985 | 568 | 2545 | 520 | . | 3633 | 204 | 3837 |
| 1986 | 858 | 1528 | 551 | . | 2937 | 995 | 3932 |
| 1987 | 1100 | 2086 | 655 | . | 3841 | 950 | 4791 |
| 1988 | 667 | 2057 | 578 | . | 3302 | 551 | 3853 |
| 1989 | 646 | 1610 | 689 | . | 2945 | 860 | 3805 |
| 1990 | 996 | 1255 | 785 | . | 3036 | 611 | 3647 |
| 1991 | 904 | 2054 | 826 | . | 3784 | 567 | 4351 |
| 1992 | 891 | 2187 | 706 | 10 | 3794 | 278 | 4072 |
| 1993 | 917 | 2322 | 610 | 13 | 3862 | 437 | 4299 |
| 1994 | 940 | 2382 | 701 | 14 | 4037 | 346 | 4383 |
| 1995 | 817 | 2248 | 669 | 9 | 3743 | 677 | 4420 |
| 1996 | 899 | 2322 | 877 | . | 4098 | 699 | 4797 |
| 1997 | 1306 | 1702 | 933 | . | 3941 | 823 | 4764 |
| 1998 | 541 | 1703 | 803 | . | 3047 | 316 | 3363 |
| 1999 | 880 | 2251 | 769 | . | 3900 | 235 | 4135 |
| 2000 | 1021 | 2190 | 621 | . | 3832 | -356 | 3476 |
| 2001 | 1313 | 2482 | 822 | . | 4617 | -592 | 4025 |
| 2002 | 1643 | 2780 | 976 | . | 5399 | -666 | 4733 |
| 2003 | 1657 | 3475 | 1114 | 1 | 6247 | -1209 | 5038 |
| 2004 | 1485 | 3070 | 1112 | . | 5667 | -841 | 4826 |
| 2005 | 1221 | 2832 | 567 | . | 4620 | -236 | 4384 |
| 2006 | 1547 | 2627 | 678 | . | 4852 | -18 | 4834 |
| 2007 | 1530 | 2981 | 801 | 1 | 5313 | -147 | 5166 |
| 2008 | 1368 | 2880 | 724 | . | 4972 | -455 | 4517 |
| 2009 | 1475 | 2886 | 754 | 6 | 5121 | 145 | 5266 |
| 2010 | 1294 | 2407 | 674 |  | 4374 | 17 | 4391 |
| 2011 | 1181 | 2283 | 686 |  | 4150 | -17 | 4133 |
| 2012 | 920 | 2475 | 623 | 0,25 | 4018 | 29 | 4047 |
| 2013 | 954 | 2865 | 605 |  | 4424 | -34 | 4390 |
| 2014 | 1495 | 2477 | 648 |  | 4620 | 0 | 4620 |
| 2015 ** | 1048 | 1856 | 468 |  | 3372 | 69 | 3441 |

* Unallocated mainly due to misreporting
** Preliminary

Table 9.2 Uptake of the national quota and the total TAC of sole Solea solea in $7 . \mathrm{d}$ in 2015 (ICES estimated landings).

|  |  | BEL |  | FRA | UK(E\&W) |
| :--- | :--- | :--- | :--- | :--- | :--- | |  | TotAL |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Landings $(\mathrm{t})$ | 1048 | 1856 | 468 | 3372 |
| TAC $(\mathrm{t})$ | 938 | 1875 | 670 | 3258 |
|  | $12 \%$ | $3 \%$ | $-30 \%$ | $-1 \%$ |
|  | over | over | under | under |

Table 9.3 Landings percentages of 7.d sole Solea solea by gear type (GNS/GTR = gill and trammel nets; TBB = beam trawls; OTB = otter trawls) in 2013-2015.

|  | LANDINGS BY GEAR | 2013 | 2014 | 2015 |
| :--- | :--- | :--- | :--- | :--- |
| GNS/GTR | $52 \%$ | $47 \%$ | $45 \%$ |  |
| TBB | $26 \%$ | $35 \%$ | $34 \%$ |  |
| OTB | $17 \%$ | $12 \%$ | $14 \%$ |  |
| OTHER | $6 \%$ | $6 \%$ | $7 \%$ |  |

Table 9.4 Summary of the Intercatch data for 7.d sole Solea solea in 2015 (imported vs. raised data; sampled vs. estimated data).

| CATCH CATEGORY | RAISEDOR IMPORTED | SAMPLED OR Estimated | CATON | PERC |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Landings | Imported Data | Sampled Distribution | 2986657 | 89 |
| Landings | Imported Data | Estimated Distribution | 379502 | 11 |
| Discards | Raised Discards | Estimated Distribution | 147021 | 58 |
| Discards | Imported Data | Estimated Distribution | 100816 | 40 |
| Discards | Imported Data | Sampled Distribution | 7244 | 3 |

Table 9.5 Discard rates of 7.d sole Solea solea by country in 2012-2015.

|  | oVERALL | BEL | FRA | UK |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Discard rate 2012 | 0,09 | 0,09 |  |  |  |
| Discard rate 2013 | 0,11 | 0,11 |  |  |  |
| Discard rate 2014 | 0,115 | 0,08 | 0,18 | 0,01 |  |
| Discard rate 2015 | 0,07 | 0,09 | 0,06 | 0,04 |  |

Table 9.6 Catch weights at age for $7 . \mathrm{d}$ sole.

| YEAR | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ |
| :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |
| AGE |  |  |  |  |
| 1 | 0,102 | 0 | 0,100 | 0,090 |
| 2 | 0,171 | 0,173 | 0,178 | 0,182 |
| 3 | 0,225 | 0,23 | 0,234 | 0,230 |
| 4 | 0,312 | 0,302 | 0,314 | 0,281 |
| 5 | 0,386 | 0,404 | 0,380 | 0,368 |
| 6 | 0,428 | 0,436 | 0,436 | 0,394 |
| 7 | 0,439 | 0,435 | 0,417 | 0,516 |
| 8 | 0,509 | 0,524 | 0,538 | 0,543 |
| 9 | 0,502 | 0,537 | 0,529 | 0,594 |
| 10 | 0,463 | 0,583 | 0,565 | 0,595 |
| +gp | 0,6729 | 0,6283 | 0,7135 | 0,8005 |


| YEAR | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 0,135 | 0,095 | 0,102 | 0,106 | 0,12 | 0,114 | 0,103 | 0,085 | 0,099 | 0,129 |
| 2 | 0,180 | 0,175 | 0,152 | 0,154 | 0,178 | 0,161 | 0,153 | 0,147 | 0,150 | 0,176 |
| 3 | 0,212 | 0,236 | 0,226 | 0,192 | 0,238 | 0,208 | 0,203 | 0,197 | 0,186 | 0,179 |
| 4 | 0,306 | 0,295 | 0,278 | 0,271 | 0,289 | 0,266 | 0,267 | 0,247 | 0,235 | 0,230 |
| 5 | 0,363 | 0,353 | 0,360 | 0,293 | 0,349 | 0,354 | 0,29 | 0,335 | 0,288 | 0,255 |
| 6 | 0,387 | 0,407 | 0,409 | 0,358 | 0,339 | 0,394 | 0,403 | 0,384 | 0,355 | 0,333 |
| 7 | 0,437 | 0,411 | 0,459 | 0,388 | 0,47 | 0,421 | 0,391 | 0,537 | 0,381 | 0,357 |
| 8 | 0,520 | 0,482 | 0,514 | 0,472 | 0,465 | 0,43 | 0,462 | 0,553 | 0,505 | 0,385 |
| 9 | 0,502 | 0,465 | 0,553 | 0,515 | 0,487 | 0,434 | 0,459 | 0,515 | 0,484 | 0,490 |
| 10 | 0,523 | 0,538 | 0,563 | 0,547 | 0,518 | 0,478 | 0,463 | 0,766 | 0,496 | 0,494 |
| +gp | 0,6015 | 0,6176 | 0,6647 | 0,7014 | 0,5621 | 0,5656 | 0,5661 | 0,6666 | 0,6156 | 0,6536 |
|  |  |  |  |  |  |  |  |  |  |  |
| YEAR | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 0,142 | 0,139 | 0,132 | 0,130 | 0,145 | 0,108 | 0,120 | 0,114 | 0,120 | 0,135 |
| 2 | 0,165 | 0,153 | 0,159 | 0,151 | 0,142 | 0,152 | 0,162 | 0,170 | 0,179 | 0,172 |
| 3 | 0,178 | 0,188 | 0,172 | 0,189 | 0,176 | 0,211 | 0,204 | 0,208 | 0,205 | 0,208 |
| 4 | 0,229 | 0,233 | 0,235 | 0,215 | 0,223 | 0,283 | 0,253 | 0,257 | 0,255 | 0,253 |
| 5 | 0,269 | 0,292 | 0,286 | 0,260 | 0,332 | 0,288 | 0,316 | 0,277 | 0,296 | 0,303 |
| 6 | 0,324 | 0,343 | 0,343 | 0,280 | 0,377 | 0,334 | 0,375 | 0,357 | 0,304 | 0,337 |
| 7 | 0,361 | 0,390 | 0,383 | 0,290 | 0,424 | 0,367 | 0,376 | 0,381 | 0,348 | 0,368 |
| 8 | 0,405 | 0,404 | 0,417 | 0,341 | 0,427 | 0,374 | 0,393 | 0,438 | 0,403 | 0,433 |
| 9 | 0,435 | 0,503 | 0,484 | 0,358 | 0,384 | 0,493 | 0,469 | 0,482 | 0,492 | 0,570 |
| 10 | 0,465 | 0,474 | 0,435 | 0,374 | 0,459 | 0,511 | 0,420 | 0,494 | 0,509 | 0,445 |


| +gp | 0,5854 | 0,6509 | 0,6162 | 0,5354 | 0,68 | 0,5445 | 0,5308 | 0,5274 | 0,525 | 0,5369 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |
| YEAR | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ | $\mathbf{2 0 1 0}$ | $\mathbf{2 0 1 1}$ | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ | $\mathbf{2 0 1 4}$ | $\mathbf{2 0 1 5}$ |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 0,139 | 0,163 | 0,148 | 0,143 | 0,124 | 0,123 | 0,173 | 0,076 | 0,017 | 0,083 |
| 2 | 0,162 | 0,190 | 0,164 | 0,177 | 0,161 | 0,161 | 0,179 | 0,150 | 0,134 | 0,144 |
| 3 | 0,192 | 0,202 | 0,201 | 0,203 | 0,195 | 0,204 | 0,204 | 0,189 | 0,175 | 0,191 |
| 4 | 0,249 | 0,227 | 0,244 | 0,260 | 0,239 | 0,252 | 0,245 | 0,245 | 0,217 | 0,228 |
| 5 | 0,284 | 0,276 | 0,262 | 0,279 | 0,287 | 0,295 | 0,288 | 0,297 | 0,267 | 0,266 |
| 6 | 0,328 | 0,294 | 0,321 | 0,358 | 0,340 | 0,326 | 0,301 | 0,315 | 0,326 | 0,280 |
| 7 | 0,353 | 0,315 | 0,435 | 0,321 | 0,342 | 0,342 | 0,377 | 0,389 | 0,344 | 0,323 |
| 8 | 0,402 | 0,378 | 0,411 | 0,464 | 0,355 | 0,399 | 0,355 | 0,446 | 0,389 | 0,345 |
| 9 | 0,457 | 0,441 | 0,377 | 0,406 | 0,512 | 0,352 | 0,467 | 0,406 | 0,403 | 0,419 |
| 10 | 0,450 | 0,439 | 0,498 | 0,476 | 0,438 | 0,441 | 0,365 | 0,441 | 0,470 | 0,357 |
| + gp | 0,557 | 0,5206 | 0,5127 | 0,6185 | 0,4504 | 0,5216 | 0,5169 | 0,5381 | 0,462 | 0,401 |

Table 9.7 Stock weights at age for $7 . \mathrm{d}$ sole.

| YEAR | 1982 | 1983 | 1984 | 1985 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 0,059 | 0,07 | 0,067 | 0,065 |  |  |  |  |  |  |
| 2 | 0,114 | 0,135 | 0,131 | 0,129 |  |  |  |  |  |  |
| 3 | 0,167 | 0,197 | 0,192 | 0,192 |  |  |  |  |  |  |
| 4 | 0,217 | 0,255 | 0,249 | 0,254 |  |  |  |  |  |  |
| 5 | 0,263 | 0,309 | 0,304 | 0,315 |  |  |  |  |  |  |
| 6 | 0,306 | 0,359 | 0,355 | 0,376 |  |  |  |  |  |  |
| 7 | 0,347 | 0,406 | 0,403 | 0,436 |  |  |  |  |  |  |
| 8 | 0,384 | 0,448 | 0,448 | 0,495 |  |  |  |  |  |  |
| 9 | 0,418 | 0,487 | 0,490 | 0,554 |  |  |  |  |  |  |
| 10 | 0,45 | 0,522 | 0,5290 | 0,6110 |  |  |  |  |  |  |
| +gp | 0,53 | 0,6008 | 0,6265 | 0,7798 |  |  |  |  |  |  |
| YEAR | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 0,070 | 0,072 | 0,050 | 0,050 | 0,05 | 0,05 | 0,05 | 0,05 | 0,050 | 0,050 |
| 2 | 0,136 | 0,139 | 0,145 | 0,113 | 0,138 | 0,138 | 0,144 | 0,13 | 0,116 | 0,126 |
| 3 | 0,198 | 0,203 | 0,223 | 0,182 | 0,232 | 0,225 | 0,199 | 0,189 | 0,161 | 0,129 |
| 4 | 0,256 | 0,262 | 0,268 | 0,269 | 0,305 | 0,279 | 0,277 | 0,246 | 0,215 | 0,220 |
| 5 | 0,309 | 0,318 | 0,365 | 0,323 | 0,4 | 0,38 | 0,305 | 0,366 | 0,273 | 0,234 |
| 6 | 0,358 | 0,370 | 0,425 | 0,335 | 0,361 | 0,384 | 0,454 | 0,377 | 0,316 | 0,333 |
| 7 | 0,403 | 0,417 | 0,477 | 0,480 | 0,476 | 0,41 | 0,405 | 0,545 | 0,368 | 0,357 |
| 8 | 0,443 | 0,461 | 0,498 | 0,504 | 0,535 | 0,449 | 0,459 | 0,56 | 0,530 | 0,330 |


| 9 | 0,480 | 0,500 | 0,572 | 0,586 | 0,571 | 0,474 | 0,43 | 0,559 | 0,461 | 0,614 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10 | 0,5120 | 0,5360 | 0,6360 | 0,5360 | 0,507 | 0,451 | 0,528 | 0,813 | 0,470 | 0,382 |
| +gp | 0,5761 | 0,6156 | 0,7498 | 0,7135 | 0,5765 | 0,6203 | 0,5269 | 0,5664 | 0,6122 | 0,6292 |
| YEAR | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 0,050 | 0,050 | 0,050 | 0,050 | 0,050 | 0,050 | 0,050 | 0,050 | 0,050 | 0,144 |
| 2 | 0,155 | 0,139 | 0,140 | 0,128 | 0,122 | 0,127 | 0,136 | 0,151 | 0,137 | 0,157 |
| 3 | 0,176 | 0,165 | 0,158 | 0,180 | 0,148 | 0,157 | 0,179 | 0,207 | 0,185 | 0,203 |
| 4 | 0,258 | 0,220 | 0,233 | 0,205 | 0,208 | 0,216 | 0,209 | 0,249 | 0,236 | 0,241 |
| 5 | 0,286 | 0,264 | 0,299 | 0,253 | 0,402 | 0,226 | 0,258 | 0,314 | 0,265 | 0,267 |
| 6 | 0,308 | 0,317 | 0,374 | 0,277 | 0,440 | 0,223 | 0,254 | 0,376 | 0,267 | 0,309 |
| 7 | 0,366 | 0,376 | 0,363 | 0,298 | 0,395 | 0,231 | 0,301 | 0,399 | 0,273 | 0,349 |
| 8 | 0,391 | 0,404 | 0,357 | 0,324 | 0,554 | 0,253 | 0,234 | 0,418 | 0,331 | 0,401 |
| 9 | 0,438 | 0,563 | 0,450 | 0,336 | 0,443 | 0,256 | 0,326 | 0,446 | 0,504 | 0,608 |
| 10 | 0,466 | 0,494 | 0,372 | 0,323 | 0,420 | 0,301 | 0,404 | 0,444 | 0,409 | 0,425 |
| +gp | 0,6304 | 0,6536 | 0,5768 | 0,5118 | 0,6822 | 0,4204 | 0,4170 | 0,5032 | 0,4501 | 0,5602 |
| YEAR | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 0,141 | 0,139 | 0,131 | 0,141 | 0,143 | 0,050 | 0,172 | 0,085 | 0,014 | 0,048 |
| 2 | 0,161 | 0,163 | 0,158 | 0,169 | 0,149 | 0,142 | 0,165 | 0,133 | 0,089 | 0,125 |
| 3 | 0,185 | 0,195 | 0,191 | 0,186 | 0,185 | 0,189 | 0,182 | 0,168 | 0,132 | 0,176 |
| 4 | 0,246 | 0,239 | 0,250 | 0,243 | 0,210 | 0,244 | 0,243 | 0,253 | 0,181 | 0,215 |
| 5 | 0,272 | 0,286 | 0,294 | 0,278 | 0,267 | 0,277 | 0,259 | 0,324 | 0,253 | 0,263 |
| 6 | 0,326 | 0,297 | 0,368 | 0,352 | 0,316 | 0,318 | 0,284 | 0,357 | 0,311 | 0,285 |
| 7 | 0,339 | 0,340 | 0,401 | 0,341 | 0,341 | 0,336 | 0,365 | 0,432 | 0,298 | 0,303 |
| 8 | 0,394 | 0,400 | 0,476 | 0,430 | 0,326 | 0,375 | 0,344 | 0,518 | 0,335 | 0,321 |
| 9 | 0,416 | 0,433 | 0,463 | 0,449 | 0,440 | 0,386 | 0,338 | 0,461 | 0,317 | 0,498 |
| 10 | 0,461 | 0,446 | 0,402 | 0,456 | 0,416 | 0,501 | 0,521 | 0,472 | 0,399 | 0,379 |
| + gp | 0,5553 | 0,5182 | 0,5663 | 0,6598 | 0,4192 | 0,5147 | 0,4788 | 0,5709 | 0,367 | 0,462 |

Table 9.8 Tuning series 1: Belgian commercial beam trawl.

|  | Effort | Age2 | Age3 | Age4 | Age 5 | Age6 | Age7 | Age8 | Age9 | Agel 10 | Agel 1 | Agel 2 | Agel 3 | Agel 4 | Agel 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 12.8 | 69.3 | 46.1 | 298.7 | 189.6 | 57.4 | 24.7 | 10.3 | 5.1 | 8.6 | 3.1 | 5.5 | 2.4 | 2.6 | 37.9 |
| 1981 | 19 | 640.7 | 161.4 | 82.1 | 312.8 | 229.6 | 44.7 | 32.9 | 33.1 | 6.9 | 9 | 18.4 | 9.3 | 0.8 | 51.9 |
| 1982 | 23.9 | 148.7 | 980.9 | 128 | 93.4 | 155.9 | 112.6 | 38.8 | 60.1 | 15.2 | 14 | 7.4 | 12.5 | 5.9 | 54.3 |
| 1983 | 23.6 | 190.4 | 373 | 818.9 | 65.5 | 54 | 81.7 | 73.2 | 23.5 | 20.2 | 27 | 5 | 1 | 7.1 | 33 |
| 1984 | 28 | 603.8 | 347.2 | 311.2 | 436 | 53.7 | 38.5 | 104.9 | 59.9 | 25.4 | 23.2 | 25.3 | 9 | 8.2 | 42.4 |
| 1985 | 25.3 | 382.9 | 612.1 | 213 | 209.1 | 260.2 | 58.2 | 34.1 | 48 | 31 | 16.9 | 19.6 | 9.2 | 7.7 | 21.3 |
| 1986 | 23.4 | 215 | 1522.3 | 675 | 233.7 | 170.6 | 194 | 30.1 | 53.1 | 64.2 | 32.6 | 12.7 | 2.6 | 43 | 29.3 |
| 1987 | 27.1 | 843.6 | 451 | 739.3 | 724.4 | 344.5 | 232.4 | 152.7 | 25.3 | 86.5 | 56 | 56.1 | 54.5 | 9.3 | 109 |
| 1988 | 38.5 | 131.6 | 990.4 | 243.3 | 362.9 | 216.7 | 111.8 | 41.8 | 73.8 | 47 | 9.8 | 22.3 | 35.8 | 8.6 | 25.3 |
| 1989 | 35.7 | 47.5 | 512.6 | 543.6 | 748 | 276.6 | 225 | 53.1 | 36.4 | 12.7 | 4.7 | 0 | 0 | 4.7 | 27 |
| 1990 | 30.3 | 1011.4 | 1375.2 | 218.1 | 366.2 | 85.3 | 198.2 | 65.5 | 39 | 22.4 | 22.2 | 25.4 | 2.8 | 24 | 18.2 |
| 1991 | 24.3 | 320.2 | 1358.6 | 710.1 | 125.6 | 283.9 | 60.6 | 56.2 | 21 | 19.8 | 22.2 | 18 | 5.6 | 0.3 | 21.4 |
| 1992 | 22 | 499.3 | 1613.7 | 523.3 | 477.7 | 36.9 | 67.9 | 28.2 | 31.7 | 11.2 | 11.4 | 6 | 5.7 | 3.2 | 16.7 |
| 1993 | 20 | 1654.5 | 1520.4 | 889.5 | 215.5 | 78.5 | 38.9 | 40.8 | 37.8 | 11.3 | 8.7 | 13.3 | 1.5 | 3 | 22.4 |
| 1994 | 22.2 | 196.9 | 1183.2 | 1598.5 | 912.9 | 201 | 160 | 39.5 | 33.8 | 46.2 | 16 | 10.2 | 14.9 | 8.8 | 18.6 |
| 1995 | 24.2 | 206.2 | 542.7 | 671.3 | 590.9 | 409.4 | 100.6 | 40.3 | 25.4 | 14.2 | 9.3 | 5 | 11.9 | 3.4 | 8 |
| 1996 | 25 | 284.1 | 975.5 | 628.7 | 560.1 | 354.3 | 316.8 | 68.3 | 77.6 | 34.2 | 26.2 | 15.8 | 10.8 | 1.1 | 4.2 |
| 1997 | 30.9 | 196 | 1282.3 | 966.1 | 500.2 | 422.3 | 301.1 | 144.7 | 56.6 | 29.3 | 25.8 | 12.1 | 12.6 | 3.4 | 1.4 |
| 1998 | 18.1 | 254.1 | 450.3 | 375.4 | 175.1 | 54.8 | 116.1 | 95.9 | 59.1 | 12.4 | 16 | 7.7 | 2.9 | 4.4 | 19.2 |
| 1999 | 21.4 | 367.7 | 1043.6 | 640.2 | 308.3 | 94.6 | 48.7 | 90.6 | 68.3 | 28.2 | 44.7 | 22.9 | 4.7 | 8.5 | 11.3 |
| 2000 | 30.5 | 569.1 | 1170.7 | 1225.1 | 239.1 | 139.4 | 68.4 | 66.6 | 74.4 | 46 | 26.9 | 7.6 | 6.6 | 0.3 | 1.9 |
| 2001 | 32.4 | 1055.5 | 1385.4 | 375 | 617.9 | 351.1 | 105.4 | 31.6 | 15.2 | 18.7 | 35.5 | 11.6 | 6.9 | 12.3 | 4.6 |
| 2002 | 33.7 | 1267.7 | 1612.6 | 804.3 | 286.3 | 122.4 | 95.7 | 45.2 | 24.8 | 28.6 | 15.8 | 13.8 | 8 | 6 | 2.6 |
| 2003 | 47.5 | 2157.2 | 1848.1 | 1368.5 | 737 | 395.3 | 191.8 | 97.9 | 15 | 47.9 | 33.5 | 30.8 | 37.9 | 0 | 1.2 |
| 2004 | 41.6 | 959.7 | 1846.2 | 778.1 | 1050.9 | 331.1 | 82.3 | 93.5 | 30.7 | 51.2 | 22 | 34.8 | 0.7 | 8.3 | 0.7 |
| 2005 | 35.8 | 1150.8 | 1156.5 | 1259.7 | 309.1 | 201.7 | 156.5 | 74.2 | 37.9 | 16.4 | 44.8 | 1.3 | 6.2 | 0.8 | 3.3 |
| 2006 | 48.8 | 1341 | 1050.9 | 1009.4 | 885.8 | 434.9 | 370.7 | 147.7 | 79.2 | 75.7 | 35.9 | 25.4 | 27.4 | 19.5 | 4.1 |
| 2007 | 57.9 | 1736.5 | 1888.6 | 808.5 | 415.2 | 550.6 | 207.8 | 258 | 117.2 | 47.6 | 36.6 | 21.5 | 9.2 | 5.5 | 31.4 |
| 2008 | 48.5 | 249.7 | 1383.2 | 1435 | 427.6 | 217.5 | 324.1 | 137.3 | 75.7 | 65.6 | 48.5 | 7.5 | 7 | 0 | 24.7 |
| 2009 | 45.3 | 1095.4 | 1185.9 | 1333.6 | 930.5 | 280.7 | 192 | 169.8 | 68.1 | 64.8 | 42.6 | 19.4 | 24.6 | 4.9 | 37.9 |
| 2010 | 35.9 | 1470.6 | 1380.4 | 442.1 | 726.2 | 492.4 | 142.6 | 66 | 137.3 | 39.5 | 76.7 | 25.5 | 17.1 | 0 | 36.4 |
| 2011 | 34.8 | 1303.1 | 2102.8 | 861.5 | 289.3 | 292.6 | 138.9 | 47.4 | 48.4 | 37.3 | 7.7 | 37.6 | 3.9 | 0 | 10.3 |
| 2012 | 31.2 | 139.6 | 1554.9 | 1147.3 | 427.5 | 178.9 | 169.4 | 172.4 | 51.5 | 6.6 | 34.6 | 14.2 | 35.6 | 19.5 | 29.8 |
| 2013 | 35.8 | 146.7 | 1633.2 | 1205.1 | 449.1 | 187.9 | 177.9 | 181.1 | 54.1 | 6.9 | 36.3 | 14.9 | 37.4 | 20.4 | 31.3 |
| 2014 | 48.1 | 210.9 | 796.8 | 1332.9 | 1423.2 | 705.4 | 227.8 | 111.4 | 123.4 | 78.1 | 8.4 | 19.6 | 16.7 | 20.3 | 24.0 |
| 2015 | 43.4 | 144.8 | 384.1 | 541.2 | 652.1 | 907.5 | 624.1 | 197.4 | 42.0 | 118.9 | 63.8 | 30.3 | 6.2 | 13.4 | 21.5 |

Table 9.9 Tuning series 2: UK(E\&W) commercial beam trawl.

|  | Effort | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Agel 10 | Agel 1 | Agel 2 | Agel 3 | Agel 4 | Agel 5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 2.79 | 30 | 144.8 | 100.5 | 28 | 28.8 | 39.4 | 1.2 | 2.4 | 5.2 | 2.5 | 2.8 | 1.5 | 1.7 | 5.3 |
| 1987 | 5.64 | 251.8 | 106 | 143.5 | 99.2 | 18.6 | 14.6 | 37.6 | 1.4 | 0.4 | 3.3 | 1.1 | 1.5 | 3.3 | 2.4 |
| 1988 | 5.09 | 112.3 | 281.3 | 56.4 | 62.9 | 39.6 | 9 | 11.5 | 16.2 | 2 | 0.2 | 4.6 | 4.9 | 0 | 0.2 |
| 1989 | 5.65 | 162.3 | 78.1 | 144.2 | 18.2 | 31.7 | 23.1 | 5.1 | 4.2 | 16.3 | 1 | 0.6 | 2.2 | 2.7 | 12.9 |
| 1990 | 7.27 | 112.6 | 327.4 | 47.7 | 66.1 | 14.1 | 15.1 | 15.1 | 4.1 | 7.4 | 22.2 | 1.9 | 0.4 | 3.4 | 7.6 |
| 1991 | 7.67 | 349 | 139.2 | 195.2 | 8.4 | 30.7 | 5.1 | 7.4 | 10.9 | 2.7 | 1.9 | 8.4 | 0.3 | 0 | 5 |
| 1992 | 8.78 | 240.1 | 516.6 | 81.3 | 167.5 | 11.1 | 20.3 | 6.4 | 14.6 | 4.9 | 2.2 | 1.5 | 3.3 | 0.1 | 2.5 |
| 1993 | 6.4 | 174.9 | 222.5 | 218.9 | 34.6 | 52.7 | 5.2 | 10.7 | 4.5 | 3 | 3.3 | 1.1 | 1.3 | 2.1 | 2.8 |
| 1994 | 5.43 | 33.6 | 260.9 | 144.1 | 113.3 | 27.5 | 45.5 | 4.4 | 10.5 | 3.2 | 4.1 | 3.7 | 2.4 | 1.6 | 9.3 |
| 1995 | 6.89 | 181.1 | 106.9 | 220.4 | 107.6 | 94.6 | 18.3 | 37.5 | 5.4 | 9.4 | 2 | 4.3 | 4.4 | 0.9 | 7.7 |
| 1996 | 10.31 | 295.8 | 251.3 | 79.5 | 169 | 84.6 | 67.4 | 17.5 | 33.2 | 4.1 | 8.8 | 4.2 | 5.4 | 3.6 | 11.9 |
| 1997 | 10.25 | 268.5 | 331.1 | 158.5 | 42.4 | 125.2 | 50.8 | 48.7 | 11.6 | 23 | 2.7 | 7.1 | 1.1 | 3.8 | 7.6 |
| 1998 | 7.31 | 252.6 | 169.4 | 97.5 | 65.2 | 22.1 | 51.7 | 28.8 | 22.4 | 5.8 | 12.5 | 2 | 5.3 | 1.5 | 9 |
| 1999 | 5.86 | 170 | 300 | 105.6 | 43.6 | 31.8 | 12.3 | 26.3 | 12.9 | 7.3 | 3.4 | 3.8 | 0.7 | 2.5 | 4.1 |
| 2000 | 5.65 | 152.1 | 178.8 | 171.4 | 54.7 | 25.8 | 18.2 | 6.9 | 21.6 | 9.7 | 5.7 | 2.3 | 4.2 | 0.6 | 7.9 |
| 2001 | 7.64 | 284.3 | 268 | 101 | 111.9 | 44 | 19 | 19.6 | 5.8 | 14.7 | 12.1 | 5 | 1.4 | 3 | 4.7 |
| 2002 | 7.9 | 314.6 | 449 | 222.2 | 71.7 | 54.9 | 22.9 | 18.6 | 6 | 3.1 | 5.2 | 2.3 | 2.4 | 0.4 | 2.9 |
| 2003 | 6.69 | 386 | 220.8 | 149.5 | 64.8 | 27.2 | 32 | 15 | 5.6 | 5.8 | 0.9 | 4.2 | 2.8 | 1.9 | 5.1 |
| 2004 | 4.87 | 111.94 | 440.41 | 103.2 | 62.24 | 32.62 | 9.61 | 18.18 | 4.33 | 3.21 | 2.89 | 0.54 | 3.32 | 1.2 | 4.22 |
| 2005 | 6 | 170.74 | 178.27 | 376.44 | 69.41 | 72.25 | 35.36 | 17.41 | 15.58 | 11.22 | 4.26 | 7.89 | 2.68 | 3.2 | 10.94 |
| 2006 | 5.94 | 395.17 | 350.51 | 113.46 | 188.96 | 31.71 | 28.12 | 13.55 | 9.03 | 5.42 | 2.76 | 0.81 | 1.49 | 0.26 | 2.92 |
| 2007 | 5 | 167.78 | 303.67 | 114.86 | 34.62 | 102.76 | 23.99 | 23.55 | 9.39 | 1.33 | 4.14 | 2.77 | 0.93 | 1.83 | 5.95 |
| 2008 | 6.21 | 152.52 | 612.94 | 184.74 | 40.66 | 24.66 | 34.21 | 12.57 | 4.41 | 6.36 | 4.55 | 1.27 | 2.28 | 0.11 | 3.56 |
| 2009 | 6.21 | 289.96 | 113.51 | 272.97 | 98.85 | 15.33 | 12.47 | 26.55 | 7.68 | 13.8 | 2.69 | 0.27 | 1.86 | 1.9 | 0.89 |
| 2010 | 4.35 | 153.05 | 151.85 | 50.86 | 101.02 | 33.93 | 11.9 | 7.8 | 14.04 | 4.89 | 3.38 | 3.7 | 0.63 | 0.57 | 2.79 |
| 2011 | 3 | 227.03 | 121.43 | 59.61 | 16.54 | 37.19 | 10.8 | 2.5 | 2.51 | 2.57 | 0.85 | 2.13 | 0.57 | 0.07 | 0.81 |
| 2012 | 3.31 | 44.70 | 323.85 | 59.64 | 34.35 | 5.88 | 15.99 | 8.54 | 1.41 | 1.42 | 3.68 | 0.68 | 0.28 | 0.88 | 1.56 |
| 2013 | 2.88 | 15.57 | 109.60 | 200.66 | 36.49 | 21.35 | 6.73 | 9.04 | 2.68 | 0.84 | 0.43 | 2.17 | 0.50 | 0.30 | 1.06 |
| 2014 | 3.02 | 75.63 | 72.96 | 164.94 | 95.63 | 14.27 | 8.56 | 1.03 | 5.96 | 2.25 | 1.18 | 0.24 | 0.66 | 0.00 | 0.52 |
| 2015 | 4.19 | 57.68 | 54.11 | 28.85 | 55.41 | 41.61 | 5.80 | 3.73 | 0.98 | 1.53 | 1.04 | 0.31 | 0.25 | 0.39 | 0.57 |

Table 9.10 Tuning series 3 : UK(E\&W)-BTS-Q3.

|  | Effort | Age 1 | Age2 | Age3 | Age4 | Age5 | Age6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 1 | 3.01 | 22.09 | 4.62 | 2.45 | 0.56 | 0.35 |
| 1990 | 1 | 17.96 | 5.55 | 5.55 | 1.24 | 1.01 | 0.33 |
| 1991 | 1 | 12.14 | 31.17 | 3.19 | 2.82 | 0.48 | 0.67 |
| 1992 | 1 | 1.33 | 15.29 | 13.47 | 1.07 | 1.61 | 0.34 |
| 1993 | 1 | 0.82 | 22.96 | 11.42 | 9.97 | 1.14 | 1.52 |
| 1994 | 1 | 8.33 | 4.26 | 11.07 | 4.65 | 4.30 | 0.28 |
| 1995 | 1 | 5.89 | 16.09 | 2.22 | 3.51 | 1.67 | 2.12 |
| 1996 | 1 | 5.30 | 10.79 | 5.97 | 1.07 | 1.86 | 1.15 |
| 1997 | 1 | 24.75 | 10.85 | 4.42 | 1.94 | 0.26 | 0.82 |
| 1998 | 1 | 3.27 | 24.11 | 3.67 | 1.47 | 0.83 | 0.19 |
| 1999 | 1 | 35.99 | 8.22 | 11.33 | 1.59 | 0.73 | 1.02 |
| 2000 | 1 | 14.98 | 27.45 | 5.52 | 4.85 | 1.48 | 0.68 |
| 2001 | 1 | 10.19 | 27.88 | 11.55 | 1.67 | 2.33 | 0.75 |
| 2002 | 1 | 53.56 | 16.11 | 8.60 | 5.11 | 0.45 | 1.04 |
| 2003 | 1 | 11.03 | 45.65 | 5.87 | 3.20 | 2.05 | 0.42 |
| 2004 | 1 | 12.67 | 11.81 | 10.97 | 2.08 | 2.02 | 1.34 |
| 2005 | 1 | 43.27 | 6.91 | 3.50 | 5.18 | 1.90 | 1.15 |
| 2006 | 1 | 10.84 | 42.62 | 4.51 | 2.68 | 2.59 | 0.55 |
| 2007 | 1 | 2.57 | 28.97 | 15.45 | 1.47 | 1.04 | 1.56 |
| 2008 | 1 | 3.77 | 7.35 | 9.14 | 5.82 | 0.40 | 0.68 |
| 2009 | 1 | 51.25 | 19.16 | 7.10 | 5.81 | 5.02 | 0.44 |
| 2010 | 1 | 16.59 | 30.76 | 5.14 | 1.66 | 2.70 | 2.73 |
| 2011 | 1 | 13.66 | 28.60 | 14.70 | 1.66 | 0.54 | 2.62 |
| 2012 | 1 | 1.75 | 9.72 | 7.51 | 3.53 | 0.92 | 0.39 |
| 2013 | 1 | 0.72 | 8.91 | 15.09 | 9.72 | 3.23 | 1.12 |
| 2014 | 1 | 25.39 | 16.35 | 12.38 | 11.92 | 5.09 | 2.73 |
| 2015 | 1 | 25.24 | 21.36 | 6.04 | 2.29 | 4.51 | 2.08 |

Table 9.11 Tuning series 4 : UK(E\&W)-YFS.

|  |  | EfFORT |
| :--- | :--- | :--- |
| 1987 | 1 | 1.38 |
| 1988 | 1 | 1.87 |
| 1989 | 1 | 0.62 |
| 1990 | 1 | 1.9 |
| 1991 | 1 | 3.69 |
| 1992 | 1 | 1.5 |
| 1993 | 1 | 1.33 |
| 1994 | 1 | 2.68 |
| 1995 | 1 | 2.91 |
| 1996 | 1 | 0.57 |
| 1997 | 1 | 1.12 |
| 1998 | 1 | 1.12 |
| 1999 | 1 | 1.47 |
| 2000 | 1 | 2.47 |
| 2001 | 1 | 0.38 |
| 2002 | 1 | 4.15 |
| 2003 | 1 | 1.44 |
| 2004 | 1 | 2.72 |
| 2005 | 1 | 4.07 |
| 2006 | 1 | 2.21 |

Table 9.12 Tuning series 5 : FR-YFS.

|  |  | Age 1 |
| :---: | :---: | :---: |
| 1987 | 1 | 0.07 |
| 1988 | 1 | 0.17 |
| 1989 | 1 | 0.14 |
| 1990 | 1 | 0.54 |
| 1991 | 1 | 0.38 |
| 1992 | 1 | 0.22 |
| 1993 | 1 | 0.03 |
| 1994 | 1 | 0.7 |
| 1995 | 1 | 0.28 |
| 1996 | 1 | 0.15 |
| 1997 | 1 | 0.03 |
| 1998 | 1 | 0.1 |
| 1999 | 1 | 0.35 |
| 2000 | 1 | 0.31 |
| 2001 | 1 | 1.21 |
| 2002 | 1 | 0.11 |
| 2003 | 1 | 0.32 |
| 2004 | 1 | 0.15 |
| 2005 | 1 | 0.82 |
| 2006 | 1 | 0.83 |
| 2007 | 1 | 0.08 |
| 2008 | 1 | 0.06 |
| 2009 | 1 | 2.78 |
| 2010 | 1 | 0.1 |
| 2011 | 1 | 0.32 |
| 2012 | 1 | 0.35 |
| 2013 | 1 | 0.05 |
| 2014 | 1 | 0.04 |
| 2015 | 1 | 0.09 |

Table 9.13 Impact of revision of UK(E\&W)-Q3-BTS: comparison of the 2015 assessment results (OLD) and a new run with the revised index (NEW).

|  | RECRUITS |  | SSB |  | FBAR3-7 |  | Y/SSB |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | OLD | NEW | OLD | NEW | OLD | NEW | OLD | NEW |
| 1986 | 25724 | 25725 | 10574 | 10562 | 0.39 | 0.39 | 0.37 | 0.37 |
| 1987 | 10979 | 10982 | 8974 | 8967 | 0.62 | 0.62 | 0.53 | 0.53 |
| 1988 | 25750 | 25731 | 10128 | 10117 | 0.44 | 0.44 | 0.38 | 0.38 |
| 1989 | 16820 | 16821 | 8415 | 8394 | 0.59 | 0.59 | 0.45 | 0.45 |
| 1990 | 44185 | 44147 | 9579 | 9568 | 0.39 | 0.39 | 0.38 | 0.38 |
| 1991 | 34859 | 34820 | 8760 | 8759 | 0.47 | 0.47 | 0.50 | 0.50 |
| 1992 | 33621 | 33580 | 11144 | 11135 | 0.38 | 0.38 | 0.37 | 0.37 |
| 1993 | 16791 | 16783 | 13108 | 13097 | 0.31 | 0.31 | 0.33 | 0.33 |
| 1994 | 26554 | 26526 | 12543 | 12528 | 0.37 | 0.37 | 0.35 | 0.35 |
| 1995 | 19403 | 19373 | 11088 | 11066 | 0.40 | 0.40 | 0.40 | 0.40 |
| 1996 | 18907 | 18886 | 12135 | 12106 | 0.50 | 0.51 | 0.40 | 0.40 |
| 1997 | 27739 | 27726 | 10511 | 10467 | 0.62 | 0.62 | 0.45 | 0.46 |
| 1998 | 17994 | 18000 | 8106 | 8069 | 0.49 | 0.49 | 0.41 | 0.42 |
| 1999 | 26250 | 26253 | 9002 | 8954 | 0.57 | 0.57 | 0.46 | 0.46 |
| 2000 | 31357 | 31356 | 8512 | 8456 | 0.45 | 0.46 | 0.41 | 0.41 |
| 2001 | 26528 | 26535 | 7600 | 7560 | 0.43 | 0.43 | 0.53 | 0.53 |
| 2002 | 46268 | 46286 | 8549 | 8532 | 0.40 | 0.41 | 0.55 | 0.56 |
| 2003 | 20950 | 20953 | 10395 | 10363 | 0.39 | 0.39 | 0.48 | 0.49 |
| 2004 | 19301 | 19308 | 11381 | 11360 | 0.40 | 0.40 | 0.42 | 0.43 |
| 2005 | 33804 | 33805 | 11416 | 11405 | 0.39 | 0.39 | 0.38 | 0.38 |
| 2006 | 40790 | 40779 | 9887 | 9888 | 0.44 | 0.44 | 0.49 | 0.49 |
| 2007 | 19899 | 19900 | 10388 | 10398 | 0.51 | 0.51 | 0.50 | 0.50 |
| 2008 | 19921 | 19899 | 12675 | 12683 | 0.43 | 0.43 | 0.36 | 0.36 |
| 2009 | 31058 | 30868 | 11540 | 11546 | 0.54 | 0.54 | 0.46 | 0.46 |
| 2010 | 40616 | 40367 | 8990 | 8995 | 0.49 | 0.49 | 0.49 | 0.49 |
| 2011 | 25227 | 24314 | 10186 | 10164 | 0.43 | 0.43 | 0.41 | 0.41 |
| 2012 | 11073 | 10381 | 12207 | 12136 | 0.42 | 0.42 | 0.33 | 0.33 |
| 2013 | 14271 | 13572 | 13384 | 13168 | 0.46 | 0.47 | 0.32 | 0.32 |
| 2014 | 25732 | 20342 | 9052 | 8772 | 0.55 | 0.58 | 0.48 | 0.50 |

Table 9.14 Impact of revision of BEL-CBT: comparison of the 2015 assessment results (OLD) and a new run with the revised index (NEW).

|  | RECRUITS |  |  | SSB |  | FBAR3-7 |  |  | Y/SSB |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | OLD | NEW | OLD | NEW | OLD |  | NEW | OLD | NEW |
| 1986 | 25724 | 25736 | 10574 | 10596 | 0.39 | 0.39 | 0.37 | 0.37 |  |
| 1987 | 10979 | 10976 | 8974 | 8988 | 0.62 | 0.62 | 0.53 | 0.53 |  |
| 1988 | 25750 | 25829 | 10128 | 10149 | 0.44 | 0.44 | 0.38 | 0.38 |  |
| 1989 | 16820 | 16826 | 8415 | 8439 | 0.59 | 0.59 | 0.45 | 0.45 |  |
| 1990 | 44185 | 44210 | 9579 | 9625 | 0.39 | 0.39 | 0.38 | 0.38 |  |
| 1991 | 34859 | 34865 | 8760 | 8790 | 0.47 | 0.47 | 0.50 | 0.49 |  |
| 1992 | 33621 | 33649 | 11144 | 11184 | 0.38 | 0.38 | 0.37 | 0.36 |  |
| 1993 | 16791 | 16783 | 13108 | 13145 | 0.31 | 0.31 | 0.33 | 0.33 |  |
| 1994 | 26554 | 26566 | 12543 | 12579 | 0.37 | 0.37 | 0.35 | 0.35 |  |
| 1995 | 19403 | 19409 | 11088 | 11120 | 0.40 | 0.40 | 0.40 | 0.40 |  |
| 1996 | 18907 | 18915 | 12135 | 12161 | 0.50 | 0.50 | 0.40 | 0.39 |  |
| 1997 | 27739 | 27768 | 10511 | 10575 | 0.62 | 0.61 | 0.45 | 0.45 |  |
| 1998 | 17994 | 17995 | 8106 | 8124 | 0.49 | 0.49 | 0.41 | 0.41 |  |
| 1999 | 26250 | 26268 | 9002 | 9028 | 0.57 | 0.57 | 0.46 | 0.46 |  |
| 2000 | 31357 | 31347 | 8512 | 8530 | 0.45 | 0.45 | 0.41 | 0.41 |  |
| 2001 | 26528 | 26536 | 7600 | 7628 | 0.43 | 0.43 | 0.53 | 0.53 |  |
| 2002 | 46268 | 46292 | 8549 | 8553 | 0.40 | 0.40 | 0.55 | 0.55 |  |
| 2003 | 20950 | 20954 | 10395 | 10416 | 0.39 | 0.39 | 0.48 | 0.48 |  |
| 2004 | 19301 | 19271 | 11381 | 11398 | 0.40 | 0.40 | 0.42 | 0.42 |  |
| 2005 | 33804 | 33760 | 11416 | 11437 | 0.39 | 0.39 | 0.38 | 0.38 |  |
| 2006 | 40790 | 41101 | 9887 | 9906 | 0.44 | 0.44 | 0.49 | 0.49 |  |
| 2007 | 19899 | 19997 | 10388 | 10385 | 0.51 | 0.51 | 0.50 | 0.50 |  |
| 2008 | 19921 | 20355 | 12675 | 12727 | 0.43 | 0.43 | 0.36 | 0.36 |  |
| 2009 | 31058 | 31266 | 11540 | 11597 | 0.54 | 0.54 | 0.46 | 0.45 |  |
| 2010 | 40616 | 42712 | 8990 | 9123 | 0.49 | 0.49 | 0.49 | 0.48 |  |
| 2011 | 25227 | 28942 | 10186 | 10377 | 0.43 | 0.42 | 0.41 | 0.40 |  |
| 2012 | 11073 | 12666 | 12207 | 12705 | 0.42 | 0.40 | 0.33 | 0.32 |  |
| 2013 | 14271 | 15954 | 13384 | 14501 | 0.46 | 0.43 | 0.32 | 0.29 |  |
| 2014 | 25732 | 25295 | 9052 | 10229 | 0.55 | 0.46 | 0.48 | 0.43 |  |
|  |  |  |  |  |  |  |  |  |  |

Table 9.15 Impact of revisions of UK(E\&W)-Q3-BTS and BEL-CBT: comparison of the 2015 assessment results (OLD) and a new run with the revised indices (NEW).

|  | RECRUITS |  | SSB |  | FBAR3-7 |  | Y/SSB |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | OLD | NEW | OLD | NEW | OLD | NEW | OLD | NEW |
| 1986 | 25724 | 25741 | 10574 | 10582 | 0.39 | 0.39 | 0.37 | 0.37 |
| 1987 | 10979 | 10982 | 8974 | 8981 | 0.62 | 0.62 | 0.53 | 0.53 |
| 1988 | 25750 | 25812 | 10128 | 10137 | 0.44 | 0.44 | 0.38 | 0.38 |
| 1989 | 16820 | 16827 | 8415 | 8418 | 0.59 | 0.59 | 0.45 | 0.45 |
| 1990 | 44185 | 44169 | 9579 | 9614 | 0.39 | 0.39 | 0.38 | 0.38 |
| 1991 | 34859 | 34820 | 8760 | 8791 | 0.47 | 0.47 | 0.50 | 0.49 |
| 1992 | 33621 | 33603 | 11144 | 11176 | 0.38 | 0.38 | 0.37 | 0.36 |
| 1993 | 16791 | 16775 | 13108 | 13140 | 0.31 | 0.31 | 0.33 | 0.33 |
| 1994 | 26554 | 26535 | 12543 | 12561 | 0.37 | 0.37 | 0.35 | 0.35 |
| 1995 | 19403 | 19377 | 11088 | 11099 | 0.40 | 0.40 | 0.40 | 0.40 |
| 1996 | 18907 | 18887 | 12135 | 12132 | 0.50 | 0.51 | 0.40 | 0.40 |
| 1997 | 27739 | 27753 | 10511 | 10528 | 0.62 | 0.62 | 0.45 | 0.45 |
| 1998 | 17994 | 18002 | 8106 | 8081 | 0.49 | 0.49 | 0.41 | 0.42 |
| 1999 | 26250 | 26271 | 9002 | 8974 | 0.57 | 0.57 | 0.46 | 0.46 |
| 2000 | 31357 | 31340 | 8512 | 8466 | 0.45 | 0.46 | 0.41 | 0.41 |
| 2001 | 26528 | 26541 | 7600 | 7583 | 0.43 | 0.43 | 0.53 | 0.53 |
| 2002 | 46268 | 46317 | 8549 | 8532 | 0.40 | 0.41 | 0.55 | 0.56 |
| 2003 | 20950 | 20958 | 10395 | 10379 | 0.39 | 0.39 | 0.48 | 0.49 |
| 2004 | 19301 | 19276 | 11381 | 11373 | 0.40 | 0.40 | 0.42 | 0.43 |
| 2005 | 33804 | 33756 | 11416 | 11422 | 0.39 | 0.39 | 0.38 | 0.38 |
| 2006 | 40790 | 41098 | 9887 | 9906 | 0.44 | 0.44 | 0.49 | 0.49 |
| 2007 | 19899 | 19999 | 10388 | 10393 | 0.51 | 0.51 | 0.50 | 0.50 |
| 2008 | 19921 | 20333 | 12675 | 12734 | 0.43 | 0.43 | 0.36 | 0.36 |
| 2009 | 31058 | 31040 | 11540 | 11600 | 0.54 | 0.54 | 0.46 | 0.45 |
| 2010 | 40616 | 42538 | 8990 | 9128 | 0.49 | 0.49 | 0.49 | 0.48 |
| 2011 | 25227 | 28205 | 10186 | 10355 | 0.43 | 0.42 | 0.41 | 0.40 |
| 2012 | 11073 | 11878 | 12207 | 12640 | 0.42 | 0.40 | 0.33 | 0.32 |
| 2013 | 14271 | 14588 | 13384 | 14315 | 0.46 | 0.43 | 0.32 | 0.30 |
| 2014 | 25732 | 20574 | 9052 | 9968 | 0.55 | 0.48 | 0.48 | 0.44 |

Table 9.16 XSA diagnostics.
FLR XSA Diagnostics 2016-06-08 12:18:43
CPUE data from indices

Catch data for 34 years. 1982 to 2015. Ages 1 to 11.
fleet first age last age first year last year alpha beta
1 BE-CBT 2101986201501

2 UK(E\&W)-CBT $2 \quad 101986201501$
3 UK(E\&W)-BTS-Q3 $1 \begin{array}{llllll} & 6 & 1989 & 2015 & 0.5 & 0.75\end{array}$

4 UK(E\&W)-YFS $11 \begin{array}{llllll} & 1987 & 2006 & 0.5 & 0.75\end{array}$

5 FR-YFS 1111198720150.50 .75

Time series weights :
Tapered time weighting not applied
Catchability analysis :
Catchability independent of size for all ages

Catchability independent of age for ages > 7

Terminal population estimation :

Survivor estimates shrunk towards the mean F of the final 5 years or the 5 oldest ages.
S.E. of the mean to which the estimates are shrunk $=2$

Minimum standard error for population estimates derived from each fleet = 0.3 prior weighting not applied

Regression weights
year
age 2006200720082009201020112012201320142015
all 11111111111

Fishing mortalities

```
year
age 2006200720082009201020112012201320142015
10.0160.009 0.008 0.008 0.004 0.000 0.000 0.001 0.003 0.001
20.2940.209 0.170 0.190 0.217 0.140 0.045 0.112 0.152 0.167
30.4160.4710.4070.4100.4190.3790.276 0.2850.4450.305
40.4830.6070.471 0.6230.5390.466 0.445 0.389 0.6230.574
50.4790.456 0.494 0.4810.535 0.450 0.377 0.420 0.4610.586
60.4140.518 0.446 0.589 0.478 0.4520.539 0.354 0.499 0.461
7 0 . 4 2 9 0 . 4 9 0 0 . 3 2 2 0 . 5 9 9 0 . 4 5 3 0 . 3 3 3 0 . 3 3 4 0 . 6 6 7 0 . 4 9 7 0 . 6 8 7 )
80.3790.4280.400 0.446 0.500 0.424 0.4820.430 1.0840.608
90.4860.3930.2310.551 0.3920.507 0.518 0.4370.5530.952
100.5660.8850.3240.5820.3430.1820.3430.676 0.460 0.653
110.5660.8850.3240.5820.343 0.1820.3430.676 0.4600.653
XSA population number (Thousand)
    age
year 1 2 3 4 5 6 7 8 91011
200640884 30438116007740948134192400 1344522506743
200720200 364152053169264320531320451414833291410
200820361 181112674011600 3415 2478 28641134 834509935
20093172118282138191610265571886143618786885991070
2010415042848313680 829878173669947 7131088 3591047
2 0 1 1 2 9 4 3 3 3 7 4 2 2 2 0 7 5 5 ~ 8 1 4 4 4 3 8 0 4 1 4 1 2 0 5 8 5 4 5 3 9 2 6 6 5 1 4 0 3 ~
20121200126632294491285146242527 238413353232131403
201398661085523045 202207452286913331546746174368
20141690289228782156741240044311822619910436664
2015 257741524369385090 76097074 24351003189474498
```

Estimated population abundance at 1st Jan 2016
age
year 14234567891011

2016123301116714627259338324037110949466223

Fleet: BE-CBT

Log catchability residuals.
year
age 1986198719881989199019911992199319941995199619971998 199920002001
$20.1140 .660-0.648-2.4841 .197-0.6850 .0481 .389-0.214-0.672-0.034-0.649-$ 0.2590 .4620 .1420 .565
$30.758-0.177-0.4000 .0320 .1240 .8630 .1290 .2870 .009-0.256-0.0150 .422$ 0.1780 .0690 .4570 .072
$40.2140 .385-0.703-0.375-0.1220 .0950 .422-0.0190 .590-0.3130 .2980 .380$ $0.3050 .5490 .363-0.319$
$5-0.0430 .634-0.1791 .054-0.0380 .0050 .2950 .0140 .316-0.017-0.0700 .515-$ $0.1000 .526-0.2580 .169$
$6-0.1270 .891-0.2310 .264-0.2040 .625-0.508-0.8570 .3950 .0640 .1150 .134-$ $0.274-0.0860 .0890 .704$
$7-0.2150 .570-0.0040 .3060 .5160 .009-0.265-0.037-0.001-0.0670 .2080 .191$ -$0.257-0.039-0.2560 .140$
$8-0.022-0.102-0.788-0.126-0.283-0.072-0.222-0.294-0.255-1.131-0.077-0.238$ $0.039-0.2410 .481-0.671$
$90.7020 .223-0.739-0.3640 .269-0.683-0.0780 .611-0.2320 .133-0.1630 .012-$ $0.089-0.030-0.292-0.648$
$100.0242 .1031 .254-2.073-0.1270 .446-0.670-0.6101 .283-0.8101 .054-0.975-$ $0.147-0.575-0.366-1.388$

## year

age 2002200320042005200620072008200920102011201220132014 2015

```
2 0.927 0.545 0.642 1.084 0.333 0.201-0.881 0.666 0.763 0.363-1.467-0.625-
0.342-1.144
3 0.122 0.164-0.314 0.111-0.191-0.321-0.749-0.173 0.225 0.243-0.348-0.187-
0.161-0.617
4-0.092 0.020-0.203-0.170 0.060-0.167 0.007-0.258-0.504 0.181 0.111-0.457-
0.2910.013
5-0.217-0.075 0.301-0.606-0.370-0.523-0.065 0.123-0.044-0.392 0.020-0.526-
0.158-0.291
6-0.842 0.439-0.121-0.562 0.034-0.295-0.316 0.344 0.424-0.198-0.048-0.347
0.3110.181
7-0.253-0.433-0.599-0.321 0.149-0.413-0.205 0.156 0.441-0.385-0.225 0.418-
0.017 0.887
8-0.341-0.190-0.554-0.102 -0.215 0.144-0.102 -0.305 -0.025 -0.090 0.441 0.183
0.5970.589
9-0.609-1.476 -0.882-0.793 0.156-0.132-0.469-0.166 0.237 0.299 0.669-0.294
0.089 0.854
100.313 0.099 0.227-0.991 0.178 0.236-0.074-0.064 0.078-0.641-1.052-0.791
0.3250.851
```

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

```
2
```

Mean_Logq -7.1528-5.8516-5.7069-5.6125-5.7380-5.6514-5.6514-5.6514-
5.6514
S.E_Logq 0.85520 .35600 .32740 .36540 .42120 .34080 .38040 .52230 .8719
Fleet: UK(E\&W)-CBT
Log catchability residuals.
year
age 1986198719881989199019911992199319941995199619971998 199920002001

```
2-0.346 0.403 0.600-0.029-0.188-0.063-0.383-0.336-1.191-0.163 0.274 0.152 \(0.0250 .368-0.1090 .080\)
3 0.506-0.081 0.338-0.032 0.090-0.288-0.117-0.522-0.120-0.650-0.511 0.145\(0.2750 .0910 .238-0.152\)
\(40.5280 .408-0.0490 .234-0.1220 .050-0.429-0.189-0.315-0.078-0.791-0.231-\) \(0.0440 .1350 .175-0.094\)
\(50.2850 .5390 .415-0.4950 .000-1.2230 .488-0.353-0.039-0.141-0.059-0.526\) 0.1410 .1880 .2760 .228
\(60.454-0.2250 .3270 .175-0.343-0.212-0.5570 .1180 .0470 .089-0.1980 .255-\) 0.0420 .3520 .3220 .305
7 0.703-0.242-0.113 0.260-0.245-0.926-0.168-0.523 0.536-0.128-0.067-0.099 0.2270 .2670 .4920 .258
\(8-0.7310 .4520 .332-0.2390 .063-0.560-0.400-0.107-0.1450 .440-0.1670 .163\) 0.1290 .2040 .2860 .682
\(90.118-0.7160 .155-0.293-0.1700 .2000 .4510 .0090 .3930 .2270 .260-0.083\) \(0.234-0.0150 .5440 .219\)
\(100.024-1.3180 .5070 .4070 .579-0.007-0.191-0.4110 .4080 .4200 .2050 .273\) \(0.387-0.2450 .1500 .202\)
```


## year

age 2002200320042005200620072008200920102011201220132014 2015
$20.3670 .1670 .0210 .3450 .600-0.3040 .0650 .707-0.0070 .449-0.980-0.965$ 0.783-0.344
$30.268-0.0260 .3720 .0010 .7910 .2740 .467-0.5580 .103-0.1840 .301-0.394$ 0.190-0.265
$40.165-0.1420 .0150 .5010 .0730 .4240 .1050 .236-0.4630 .053-0.5100 .3630 .480$ -0.488

5 0.172-0.224-0.058 0.009 0.514-0.235-0.039 0.191 0.417-0.480 0.065-0.193 0.233-0.096
$60.041-0.043-0.0600 .431-0.2450 .710-0.204-0.3420 .0930 .424-0.9860 .232-$ 0.587-0.330

7 0.154 0.122-0.215 0.364 0.062 0.263-0.012-0.205 0.454-0.102 0.045 0.050-0.144-1.067

```
8 0.608 0.280 0.340 0.620-0.112 0.585-0.051 0.213 0.337-0.195 0.066 0.092-
0.932-0.656
9-0.191-0.115-0.310 0.490 0.477 0.179-0.870 0.025 0.454 0.177-0.299-0.393
0.213-0.179
10-0.072 0.335-0.011 0.802 0.033-0.506 0.034 0.763 0.486-0.479 0.041 0.010-
0.068-0.778
```

Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time
$\begin{array}{lllllllll}2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10\end{array}$

Mean_Logq -6.5355-5.8256-5.7995-5.9355-5.9716-6.0377-6.0377-6.03776.0377
S.E_Logq 0.48080 .34620 .33280 .37520 .36670 .38580 .41520 .34750 .4541

Fleet: UK(E\&W)-BTS-Q3

Log catchability residuals.
year
age 1989199019911992199319941995199619971998199920002001 2002200320042005
$1-0.6890 .141-0.022-2.204-1.990-0.132-0.137-0.2450 .910-0.6811 .3430 .294$ 0.0741 .1790 .3960 .6421 .275
$20.244-0.6710 .101-0.432$ 0.029-1.046-0.123-0.175-0.204 0.185-0.347 0.447 0.3400 .028 0.493-0.094-0.523
$30.551-0.371-0.3770 .040-0.0150 .034-0.956-0.258-0.169-0.4610 .2390 .154$ $0.3780 .031-0.114-0.104-0.505$
$4-0.0630 .224-0.129-0.5850 .579-0.041-0.283-0.741-0.238-0.216-0.2510 .364-$ $0.1730 .474-0.089-0.3100 .014$
$5-0.167-0.171-0.0220 .0450 .1020 .405-0.350-0.200-1.202-0.185-0.0850 .417$ $0.438-0.7960 .2430 .1260 .217$
$6-0.658-0.201-0.0480 .0510 .328-0.9300 .153-0.219-0.505-0.8610 .6440 .348$
$0.2050 .050-0.3810 .2630 .010$
year
age 2006200720082009201020112012201320142015

1 -0.292-1.031-0.656 1.510 $0.1110 .258-0.900-1.5921 .4351 .005$
$20.7090 .091-0.6060 .3550 .4010 .007-0.7910 .0620 .8890 .631$
$3-0.1450 .549-0.2790 .130-0.1780 .432-0.6540 .2951 .1610 .592$
$40.123-0.2890 .4860 .251-0.391-0.417-0.1320 .3930 .9970 .442$
$50.040-0.101-0.7981 .0710 .309-0.774-0.3410 .4640 .4360 .881$
$6-0.5830 .083-0.029-0.1020 .9890 .810-0.5460 .2660 .8130 .049$

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time
$\begin{array}{llllll}1 & 2 & 3 & 4 & 5 & 6\end{array}$

Mean_Logq-7.8708-7.0345-7.3856-7.7281-7.8839-7.8315
S.E_Logq 1.01010 .47510 .44550 .39820 .51120 .4938

Fleet: UK(E\&W)-YFS

Log catchability residuals.
year
age 1987198819891990199119921993199419951996199719981999 20002001200220032004
$10.6410 .094-0.578-0.4150 .477-0.3930 .1840 .4240 .848-0.784-0.495-0.062-$ $0.1650 .182-1.5250 .3120 .050 .794$
year
age 20052006
$10.602-0.192$

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

1

Mean_Logq -9.5614
S.E_Logq 0.5902

Fleet: FR-YFS

Log catchability residuals.
year
age 1987198819891990199119921993199419951996199719981999 20002001200220032004

1 -0.254-0.218 $0.020 .4130 .29-0.227-1.5221 .1670 .593-0.033-2.029-0.3920 .486$
0.192 1.719-1.233 0.632-0.018
year
age 20052006200720082009201020112012201320142015
$11.0850 .915-0.724-1.022 .372-1.2240 .281 .267-0.443-1.242-0.855$

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

1

Mean_Logq -11.6473
S.E_Logq 1.0130

Terminal year survivor and $F$ summaries:

Age 1 Year class =2014
source
scaledWts survivors yrcls

UK(E\&W)-BTS-Q3 0.442636652014

```
FR-YFS 0.441 9912 2014
fshk 0.117 13068 2014
```

Age 2 Year class $=2013$
source
scaledWts survivors yrcls
BE-CBT $\quad 0.13137192013$

UK(E\&W)-CBT $0.415 \quad 82772013$

UK(E\&W)-BTS-Q3 0.424219312013
fshk 0.029149112013

Age 3 Year class $=2012$
source
scaledWts survivors yrcls
BE-CBT 0.36524972012
UK(E\&W)-CBT 0.38635502012
UK(E\&W)-BTS-Q3 0.23283642012
fshk 0.01637832012

Age 4 Year class $=2011$
source
scaledWts survivors yrcls
BE-CBT $0.372 \quad 26282011$

UK(E\&W)-CBT 0.36015922011
UK(E\&W)-BTS-Q3 0.25040342011
fshk 0.01831502011

```
Age 5 Year class =2010
source
    scaledWts survivors yrcls
BE-CBT 0.397 28632010
UK(E&W)-CBT 0.376 3481 2010
UK(E&W)-BTS-Q3 0.202 92432010
fshk 0.025 53692010
Age 6 Year class =2009
source
    scaledWts survivors yrcls
BE-CBT 0.321 4840 2009
UK(E&W)-CBT 0.423 29042009
UK(E&W)-BTS-Q3 0.233 4242 2009
fshk 0.023 39812009
```

```
Age 7 Year class =2008
source
    scaledWts survivors yrcls
BE-CBT 0.543 26922008
UK(E&W)-CBT 0.424 3812008
fshk 0.032 1883 2008
```

    Age 8 Year class \(=2007\)
    source
scaledWts survivors yrcls
BE-CBT 0.5058902007

```
UK(E&W)-CBT 0.457 256 2007
fshk 0.038 518 2007
```

Age 9 Year class $=2006$
source
scaledWts survivors yrcls
BE-CBT $0.280 \quad 1562006$
UK(E\&W)-CBT 0.665552006
fshk 0.0541702006

Age 10 Year class $=2005$
source
scaledWts survivors yrcls

BE-CBT 0.1995232005

UK(E\&W)-CBT 0.7261032005
fshk 0.0762192005

Table 9.17 Sole 7.d. XSA summary

|  | RECRUITS | SSB | CATCH | LANDINGS | FBAR3-7 | Y/SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 12686 | 7732 | 3190 | 3190 | 0.34 | 0.41 |
| 1983 | 21296 | 9532 | 3458 | 3458 | 0.38 | 0.36 |
| 1984 | 21545 | 8957 | 3575 | 3575 | 0.47 | 0.4 |
| 1985 | 12943 | 9985 | 3837 | 3837 | 0.34 | 0.38 |
| 1986 | 25756 | 10623 | 3932 | 3932 | 0.39 | 0.37 |
| 1987 | 10993 | 9016 | 4791 | 4791 | 0.62 | 0.53 |
| 1988 | 25806 | 10136 | 3853 | 3853 | 0.44 | 0.38 |
| 1989 | 16819 | 8404 | 3805 | 3805 | 0.59 | 0.45 |
| 1990 | 44324 | 9619 | 3647 | 3647 | 0.39 | 0.38 |
| 1991 | 34875 | 8837 | 4351 | 4351 | 0.47 | 0.49 |
| 1992 | 33667 | 11242 | 4072 | 4072 | 0.38 | 0.36 |
| 1993 | 16788 | 13228 | 4299 | 4299 | 0.31 | 0.32 |
| 1994 | 26539 | 12620 | 4383 | 4383 | 0.37 | 0.35 |
| 1995 | 19397 | 11164 | 4420 | 4420 | 0.4 | 0.4 |
| 1996 | 18880 | 12200 | 4797 | 4797 | 0.5 | 0.39 |
| 1997 | 27805 | 10586 | 4764 | 4764 | 0.61 | 0.45 |
| 1998 | 18041 | 8139 | 3363 | 3363 | 0.49 | 0.41 |
| 1999 | 26314 | 9081 | 4135 | 4135 | 0.57 | 0.46 |
| 2000 | 31229 | 8532 | 3476 | 3476 | 0.45 | 0.41 |
| 2001 | 26510 | 7648 | 4025 | 4025 | 0.43 | 0.53 |
| 2002 | 46410 | 8545 | 4733 | 4733 | 0.4 | 0.55 |
| 2003 | 20955 | 10385 | 5038 | 5038 | 0.39 | 0.49 |
| 2004 | 19274 | 11390 | 4826 | 4826 | 0.4 | 0.42 |
| 2005 | 33856 | 11439 | 4383 | 4383 | 0.39 | 0.38 |
| 2006 | 40884 | 9950 | 4833 | 4833 | 0.44 | 0.49 |
| 2007 | 20200 | 10436 | 5166 | 5166 | 0.51 | 0.5 |
| 2008 | 20361 | 12732 | 4517 | 4517 | 0.43 | 0.35 |
| 2009 | 31721 | 11555 | 5266 | 5266 | 0.54 | 0.46 |
| 2010 | 41504 | 9142 | 4409 | 4409 | 0.48 | 0.48 |
| 2011 | 29433 | 10542 | 4133 | 4133 | 0.42 | 0.39 |
| 2012 | 12001 | 12619 | 4048 | 4048 | 0.39 | 0.32 |
| 2013 | 9866 | 14439 | 4390 | 4390 | 0.42 | 0.3 |
| 2014 | 16902 | 10017 | 4620 | 4620 | 0.51 | 0.46 |
| 2015 | 25774 | 7899 | 3441 | 3441 | 0.52 | 0.44 |

Table 9.18 Sole 7.d. RCT3-input for Age 1.

|  | 7.d | Age1 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 35 | 2 |  |  |  |  |
| 1981 | 12686 |  | 3.33 | 0.07 | -11 | -11 |
| 1982 | 21296 |  | 1.04 | 0.02 | -11 | -11 |
| 1983 | 21545 |  | 0.79 | -11 | -11 | -11 |
| 1984 | 12943 |  | -11 | -11 | -11 | -11 |
| 1985 | 25756 |  | -11 | -11 | -11 | -11 |
| 1986 | 10993 |  | -11 | 0.07 | -11 | 14.20 |
| 1987 | 25806 |  | 0.75 | 0.17 | 8.20 | 22.09 |
| 1988 | 16819 |  | 0.04 | 0.14 | 3.01 | 5.55 |
| 1989 | 44324 |  | 17.43 | 0.54 | 17.96 | 31.17 |
| 1990 | 34875 |  | 0.57 | 0.38 | 12.14 | 15.29 |
| 1991 | 33667 |  | 1.04 | 0.22 | 1.33 | 22.96 |
| 1992 | 16788 |  | 0.48 | 0.03 | 0.82 | 4.26 |
| 1993 | 26539 |  | 0.27 | 0.70 | 8.33 | 16.09 |
| 1994 | 19397 |  | 4.04 | 0.28 | 5.89 | 10.79 |
| 1995 | 18880 |  | 3.50 | 0.15 | 5.30 | 10.85 |
| 1996 | 27805 |  | 0.28 | 0.03 | 24.75 | 24.11 |
| 1997 | 18041 |  | 0.07 | 0.10 | 3.27 | 8.22 |
| 1998 | 26314 |  | 10.52 | 0.35 | 35.99 | 27.45 |
| 1999 | 31229 |  | 2.84 | 0.31 | 14.98 | 27.88 |
| 2000 | 26510 |  | 2.41 | 1.21 | 10.19 | 16.11 |
| 2001 | 46410 |  | 4.32 | 0.11 | 53.56 | 45.65 |
| 2002 | 20955 |  | 0.94 | 0.32 | 11.03 | 11.81 |
| 2003 | 19274 |  | 0.21 | 0.15 | 12.67 | 6.91 |
| 2004 | 33856 |  | 7.29 | 0.82 | 43.27 | 42.62 |
| 2005 | 40884 |  | 0.05 | 0.83 | 10.84 | 28.97 |
| 2006 | 20200 |  | 1.04 | 0.08 | 2.57 | 7.35 |
| 2007 | 20361 |  | 0.03 | 0.06 | 3.77 | 19.16 |
| 2008 | 31721 |  | 6.58 | 2.78 | 51.25 | 30.76 |
| 2009 | 41504 |  | 2.47 | 0.10 | 16.59 | 28.60 |
| 2010 | 29433 |  | 0.20 | 0.32 | 13.66 | 9.72 |
| 2011 | 12001 |  | 2.78 | 0.35 | 1.75 | 8.91 |
| 2012 | -11 | 0.44 | 0.052 | 0.72 | 16.35 |  |
| 2013 |  | 0.72 | 0.04 | 25.39 | 21.36 |  |
| 2014 |  | 1.08 | 0.09 | 25.24 | -11 |  |
| 2015 | -11 | 0.26 | -11 | -11 | -11 |  |
| FRYF0 |  |  |  |  |  |  |
| FRYF1 |  |  |  |  |  |  |
| BTS1 |  |  |  |  |  |  |
| BTS2 |  |  |  |  |  |  |

Table 9.19 Sole 7.d. RCT3-input for Age 2.


Table 9.20 Sole 7.d. RCT3-output for Age 1.

Analysis by RCT3 ver4.0

Data for 4 surveys over 35 years : 1981-2015
Regression type $=C$

Tapered time weighting not applied

Survey weighting not applied

Final estimates not shrunk towards mean

Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as . 00

Minimum of 3 points used for regression

Forecast/Hindcast variance correction used.
yearclass:2012
index slope intercept se rsquare n indices prediction se.pred WAP.weights

FRYFO $0.950910 .1761 .62780 .0477328-0.82109 .3961 .72520 .01658$

FRYF1 $0.800711 .3940 .87140 .1713028-2.95659 .0270 .94140 .05569$

BTS1 0.4715 $9.1310 .40730 .4334925-0.32858 .9760 .47430 .21935$

BTS2 0.78937 .9410 .34080 .56507262 .794210 .1460 .36150 .37774

VPA Mean NA NA NA NA 31 NA 10.074 0.3864 0.33064
yearclass:2013
index slope intercept se rsquare n indices prediction se.pred WAP.weights

FRYFO $0.950910 .1761 .62780 .0477328-0.32859 .8641 .71990 .01618$

FRYF1 $0.800711 .3940 .87140 .1713028-3.21898 .8170 .95040 .05300$

BTS1 0.4715 9.131 0.4073 0.43349 253.234410 .6560 .44060 .24658

BTS2 0.78937 .9410 .34080 .56507263 .061510 .3570 .36290 .36352

```
VPA Mean NA NA NA NA 31 NA 10.074 0.3864 0.32072
```

yearclass:2014
index slope intercept se rsquare n indices prediction se.pred WAP.weights FRYFO 0.950910 .1761 .62780 .04773280 .0769610 .2501 .71930 .02534 FRYF1 $0.800711 .3940 .87140 .1713028-2.407959 .4660 .92770 .08702$ BTS1 0.47159 .1310 .40730 .43349253 .2284310 .6530 .44050 .38590 BTS2 0.78937 .9410 .34080 .5650726 NA NA NA NA

VPA Mean NA NA NA NA 31 NA 10.074 0.3864 0.50174
yearclass:2015
index slope intercept se rsquare n indices prediction se.pred WAP.weights

FRYFO 0.950910 .1761 .62780 .0477328 -1.347 8.8951 .73630 .04718

FRYF1 0.8007 11.394 0.8714 0.17130 28 NA NA NA NA

BTS1 0.4715 9.131 0.4073 0.43349 25 NA NA NA NA

BTS2 0.7893 7.941 0.3408 0.5650726 NA NA NA NA

VPA Mean NA NA NA NA 31 NA 10.074 0.3864 0.95282

WAP logWAP int.se
yearclass:2012 178719.7910 .2222
yearclass:2013 2829610.2500 .2188
yearclass:2014 2825910.2490 .2737
yearclass:2015 2243910.0190 .3771

Table 9.21 Sole 7.d. RCT3-output for Age 2.
Analysis by RCT3 ver4.0

Data for 4 surveys over 35 years : 1981-2015

```
Regression type = C
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as .00
Minimum of 3 points used for regression
```

Forecast/Hindcast variance correction used.
yearclass:2013
index slope intercept se rsquare n indices prediction se.pred WAP.weights
FRYFO $0.965010 .0671 .65310 .0464228-0.3285$ 9.750 1.7466 0.01565
FRYF1 $0.808711 .2960 .88200 .1676628-3.21898 .6930 .96190 .05159$
BTS1 0.47719 .0080 .41530 .42479253 .234410 .5510 .44930 .23648
BTS2 0.78297 .8480 .33480 .57323263 .061510 .2450 .35660 .37543
VPA Mean NA NA NA NA 31 NA 9.965 0.38570 .32086
yearclass:2014
index slope intercept se rsquare n indices prediction se.pred WAP.weights
FRYFO 0.965010 .0671 .65310 .04642280 .0769610 .1411 .74610 .02496
FRYF1 0.808711 .2960 .88200 .1676628 -2.40795 9.3490 .93900 .08632
BTS1 0.47719 .0080 .41530 .42479253 .2284310 .5480 .44920 .37717

BTS2 0.78297 .8480 .33480 .5732326 NA NA NA NA

VPA Mean NA NA NA NA 31 NA 9.965 0.3857 0.51155
index slope intercept se rsquare n indices prediction se.pred WAP.weights FRYFO $0.965010 .0671 .65310 .0464228-1.347 \quad 8.7671 .76330 .04567$

FRYF1 0.808711 .2960 .88200 .1676628 NA NA NA NA

BTS1 0.4771 9.008 0.4153 0.42479 25 NA NA NA NA

BTS2 0.7829 7.848 0.3348 0.57323 26 NA NA NA NA

VPA Mean NA NA NA NA 31 NA 9.965 0.3857 0.95433

WAP logWAP int.se
yearclass:2013 2532210.140 .2185
yearclass:2014 2523410.140 .2759
yearclass:2015 201289.910 .3768

Table 9.22 Sole 7.d. Short-term forecast. Management options under the TAC-constraint scenario.

| BASIS | LANDINGS | F3-7 | SSB2017 | SSB2018 | SSB_CHANGE | TAC_CHANGE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ftar | 2294 | 0.3 | 7853 | 9400 | 20 | -22 |
| Fmsy | 2257 | 0.294 | 7853 | 9440 | 20 | -24 |
| Fmsy_low | 1304 | 0.16 | 7853 | 10472 | 33 | -56 |
| Fmsy_high | 3503 | 0.5 | 7853 | 8093 | 3 | 18 |
| Fpa | 2926 | 0.4 | 7853 | 8716 | 11 | -1 |
| Flim | 3773 | 0.55 | 7853 | 7803 | -1 | 28 |
| SSB>Bpa | 3590 | 0.516 | 7853 | 8000 | 2 | 21 |
| TACsq | 2957 | 0.405 | 7853 | 8683 | 11 | 0 |
| 15\%_TAC_inc | 3401 | 0.482 | 7853 | 8204 | 4 | 15 |
| 15\%_TAC_dec | 2513 | 0.334 | 7853 | 9162 | 17 | -15 |
| Fsq* 0 | 0 | 0 | 7853 | 11887 | 51 | -100 |
| Fsq* 0.25 | 933 | 0.112 | 7853 | 10874 | 38 | -68 |
| Fsq* 0.5 | 1765 | 0.223 | 7853 | 9972 | 27 | -40 |
| Fsq**.9 | 2932 | 0.401 | 7853 | 8710 | 11 | -1 |
| Fsq* ${ }^{*}$ | 3198 | 0.446 | 7853 | 8423 | 7 | 8 |
| Fsq* 1.1 | 3453 | 0.491 | 7853 | 8147 | 4 | 17 |
| Fsq* ${ }^{*} .25$ | 3815 | 0.558 | 7853 | 7757 | -1 | 29 |
| Fsq* ${ }^{*} .5$ | 4368 | 0.669 | 7853 | 7162 | -9 | 48 |
| Fsq* ${ }^{*} .75$ | 4869 | 0.78 | 7853 | 6623 | -16 | 65 |
| Fsq*2 | 5329 | 0.892 | 7853 | 6131 | -22 | 80 |



Figure 9.1 Sole 7.d. Official landings (tonnes) by country over the period 1974-2015, as officially reported to ICES.


Figure 9.2 Relative contribution to the official landings of sole Solea solea from Division 7.d for the main countries involved over the period 1974-2015.


Figure 9.3 Uptake of the national quota and the total TAC of sole Solea solea in 7.d in 2015 (ICES landings as uploaded to InterCatch).


Figure 9.4 Overview of $20157 . \mathrm{d}$ sole Solea solea landings (and corresponding percentages) by métier and country, for which biological data were uploaded (8 left blocks; totaling to $87 \%$ ) or not available (rest of plot) in InterCatch.


Figure 9.5 Overview of 2015 7.d sole Solea solea landings (and corresponding percentages) by country, for which biological data were uploaded (3 left blocks; totaling to $87 \%$ ) or not available (rest of plot) in InterCatch.


Figure 9.6 Catch proportions at age for 7.d sole.


Figure 9.7 Standardised catch proportions at age for 7.d sole.


Figure 9.8 Catch weights at age for $7 . \mathrm{d}$ sole.


Figure 9.9 Stock weights at age for $7 . \mathrm{d}$ sole.


Figure 9.10 Impact of revision of UK(E\&W)-Q3-BTS : catchability residuals for all tuning fleets used in the assessment of $7 . d$ sole.


Figure 9.11 Impact of revision of UK(E\&W)-Q3-BTS : assessment summary.


Figure 9.12 Impact of revision of UK(E\&W)-Q3-BTS : retrospective pattern.


Figure 9.13 Impact of revision of BEL-CBT : catchability residuals for all tuning fleets used in the assessment of $7 . \mathrm{d}$ sole.


Figure 9.14 Impact of revision of BEL-CBT : assessment summary.


Figure 9.15 Impact of revision of BEL-CBT : retrospective pattern.


Figure 9.16 Impact of revisions of UK(E\&W)-Q3-BTS AND BEL-CBT : catchability residuals for all tuning fleets used in the assessment of $7 . d$ sole.


Figure 9.17 Impact of revisions of UK(E\&W)-Q3-BTS AND BEL-CBT : assessment summary.


Figure 9.18 Impact of revisions of UK(E\&W)-Q3-BTS AND BEL-CBT: retrospective pattern.


Figure 9.19 Standardised tuning indices at age for $7 . \mathrm{d}$ sole.


Figure 9.20 Internal consistency plot of the BEL-CBT tuning series for 7.d sole.

UK(E\&W)-CBT

$\log _{10}$ (Index Value)
Lower right panels show the Coefficient of Determination $\left(r^{2}\right)$

Figure 9.21 Internal consistency plot of the UK(E\&W)-CBT tuning series for 7.d sole.

UK(E\&W)-BTS-Q3


Figure 9.22 Internal consistency plot of the UK(E\&W)-BTS-Q3 tuning series for 7.d sole.


Figure 9.23 Catchability residuals for all tuning fleets used in the 2016 assessment of $7 . \mathrm{d}$ sole.


Figure 9.24 Sole 7.d. F at age.


Figure 9.25 Sole 7.d. F per cohort.


Figure 9.26 Sole 7.d. XSA Summary: trends in recruitment (rec), spawning stock biomass (SSB), fbar and landings.


Figure 9.27 Sole 7.d. Restrospective patterns in F, recruitment and SSB.

Relative contribution of yearclasse


Figure 9.28 Sole 7.d. Relative contribution of year classes to landings in 2017.

Relative contribution of yearclasse


Figure 9.29 Sole 7.d. Relative contribution of year classes to SSB in 2018.

## 10 Sole (Solea solea) in Subarea 4 (North Sea)

The assessment of sole in Subarea 4 is presented as an update assessment. The most recent benchmark assessment was carried out in February 2015 (ICES WKNSEA 2015). More details can be found in the most recent Stock Annex. Only a comprehensive description of the methods and deviations from the stock annex are presented within this Section of the report.

### 10.1 General

### 10.1.1 Stock definition

See Stock Annex.

### 10.1.2 Ecosystem aspects

No new information on ecosystem aspects was presented at WGNSSK (2016). All available information on ecological aspects can be found in the Stock Annex.

### 10.1.3 Fisheries

See Stock Annex for a general comprehensive description of the fishery.
Many vessels in the beam trawl fleet, that is mainly catching sole in the North Sea, have adopted technological developments to their gears. The catch composition of these "advanced" gears are different from the traditional beam trawl (van Marlen et al., 2014). The operational use of these new gears cannot be distinguished using current logbook data.

### 10.1.4 ICES Advice

The information in this section is taken from the update advice from section 6.3.46 in the Advice summary sheet 2015.

## Advice for 2016

ICES advises that when the second stage of the EU management plan (Council Regulation No. 676/2007) is applied, catches in 2016 should be no more than 13031 tonnes. If discard rates do not change from the average (2012-2014), this implies landings of no more than 12066 tonnes.

## Management plan

An evaluation of the management plan (ICES, 2010) concluded that the management plan is precautionary. The stocks are in stage two of the EU multiannual plan (EU, 2007). Application of stage two of the plan is based on transitional arrangements until an evaluation of the plan has been conducted. ICES assumes that harvesting the stock with the newest estimate of FMSY is in accordance with stage two of the current plan.

### 10.1.5 Management

A multiannual plan for plaice and sole in the North Sea was adopted by the EU Council in 2007 (EC regulation 676/2007) describing two stages; of which the first stage should be deemed a recovery plan and its second stage a management plan.

The plan was implemented in 2007. ICES has evaluated the plan and found it to be in agreement with the precautionary approach (ICES, 2010). A subsequent evaluation in 2012 (Coers et al., 2012) addressed amendments to the plan in the context of moving towards stage two of the plan.
As of December 2014, the management plan has officially moved to the stage two (EU, 2014).

## Mixed fishery advice

The information in this section is taken from the North Sea Advice overview section 6.3 in the ICES Advisory report 2008. The information has not been updated in 2009 2016.

Demersal fisheries in the area are mixed fisheries, with many stocks exploited together in various combinations in the various fisheries. In these cases, management advice must consider both the state of individual stocks and their simultaneous exploitation in demersal fisheries. Stocks in the poorest condition, particularly those which suffer from reduced reproductive capacity, become the overriding concern for the management of mixed fisheries, where these stocks are exploited either as a targeted species or as a by-catch. The exploitation of sole and plaice are closely connected as they are caught together in fisheries mainly targeting sole, which are more valuable. This means that the minimum mesh size is decided on the basis of the more valuable species (sole), resulting in substantial discards of undersized plaice. The mixed fisheries for flatfish are dominated by a mixed beam trawl fishery using 80 mm mesh in the southern North Sea where up to $80 \%$ in number of all plaice caught are being discarded. Additionally, a shift in the age and size at maturation of plaice has been observed (Grift et al., 2004): plaice become mature at younger ages and at smaller sizes in recent years than in the past. There is a risk that this is caused by a genetic fisheries-induced change: Those fish that are genetically programmed to mature late at large sizes are likely to have been removed from the population before they have had a chance to reproduce and pass on their genes. This shift in maturation also leads to mature fish being of a smaller size-at-age. Measures to reduce discarding in the mixed beam trawl fishery would greatly benefit the plaice stock and future yields. In order to improve the selection pattern, mesh size increases or configuration changes (i.e. square mesh) would help reduce the discards. However, this would result in a short-term loss of marketable sole. Readjustment of minimum landing sizes corresponding to an improved selection pattern could be considered.
Improvements to gear selectivity, which would contribute to a reduction in catches of small fish, must take into account the effect on the other species within the mixed fishery. For instance, mesh enlargement in the flatfish fishery would reduce the catch of undersized plaice, but would also result in loss of marketable sole.

### 10.2 Data available

### 10.2.1 Landings

Annual landings by country and TACs are presented next to the landings submitted to InterCatch in Table 10.2.1. The TAC of 11900 t in 2015 was fully taken and slightly overshot compared to official landings of 12203 t , and landings reported to ICES of 12 867 t . Landings in numbers by age that are input for the assessment model are presented in Table 10.2.2. A time series of total landings is shown on Figure 10.2.1.

### 10.2.2 Discards

Discards were included in the assessment after the most recent benchmark (WKNSEA, 2015). A time series from national discard monitoring programmes from 2002 onwards
is used since then. Discards in numbers by age from 2002 until present are shown in Table 10.2.3. A time series of total discards is shown on Figure 10.2.2.

### 10.2.3 InterCatch

Since 2012, InterCatch is used for raising the catch. Age distributions were provided by Denmark, Germany, UK, and the Netherlands, accounting for $86 \%$ of the landings in 2105 (Figure 10.2.3).

Discards estimates for 2015 were available from Belgium, Denmark, Germany, the Netherlands, and the UK (Figure 10.2.3.) for $88 \%$ of the landings weight. This implies that $78 \%$ of the discards were imported and $22 \%$ was raised.

First metiers for which yearly discard estimates had been imported were grouped with the same metiers with quarterly landings estimates. Then, discards were raised by grouping metiers with small meshes apart from metiers with larger mesh sizes, and by grouping static gears apart from towed gears. In the towed gear group a distinction was made between otter trawlers and seines, and beam trawlers. Beam trawlers and otter trawlers targeting crustaceans (CRU) with a mesh size smaller than 99 mm were grouped together. The remainder, which consisted of metiers which did not fit in any of the above groups or, were then raised with all available discard estimates.

Discard estimates from Scotland (UK) were not included in any of the groups for raising unsampled strata since they held strangely high discard ratios.
Allocation scheme to raise discards and age structures to unsampled fleets

| UnSAMPLED FLEET* | SAMPLED FLEET** |
| :--- | :--- |
| Quarterly landings | Yearly discards |
| GN/GTR/GNS | GN/GTR/GNS |
| OTB $70-99$ | OTB 70 -99 |
| OTB $100->120$ | OTB $100->120$ |
| TBB $70-99$ | TBB $70-99$ |
| TBB $100->120$ | TBB $100->120$ |
| OTB \& TBB CRU <99 | OTB \& TBB CRU <99 |
| Others | All métiers, excluding métiers for crustaceans (_CRU) |
| $*$ Unsampled fleet are those fleets for which no discards or age structure is known. |  |
| $* *$ Sampled fleet are those fleets for which the discard rate or age structure is known. |  |

### 10.2.4 Age compositions

In 2015, the age compositions of landings and discards were raised in and exported from InterCatch. The age composition of the landings and discards is presented in numbers in Table 10.2.2-3., and Figure 10.2.4.

For metiers where no age was available, age compositions were allocated using the same method as for the discard raising (described above). These allocations were done separately for discards and landings.

Both catch categories were separately exported from InterCatch. The SOP correction for the landings was 1.000 and was 1.008 for discards.

### 10.2.5 Weight at age

Since 2012 weights at age in the landings for both sexes combined (Table 10.2.4) are measured weights from the various national market sampling programmes. Discard
weights at age (Table 10.2.5) are derived from the various national discard programmes (observer and self-sampling).
Mean stock weights at age (Table 10.2.6.) are the average weights from the 2nd Quarter landings and discards and are derived from the InterCatch (CatchAndSampleDataTable).
Landing, discard, and mean stock weights at age are presented on Figure 10.2.5.
At WGNSSK 2016 the mean weights in quarter 2 for 2015 seemed to be on the low side and sharply contrasted with mean stock weights before 2012. It seems that some countries had submitted strangely low weights for age 1 and 2 . After revision of the weight sample data in quarter two, the sample data with unrealistically low weights were taken out of the InterCatch final datafile.

However, stock weights of younger ages after 2012 are still slightly lower than stock weights before 2012. This is because before deriving the mean stock weights from InterCatch (since 2012), these weights were manually raised based on landings only. In that time series (1957-2011) a constant value (0.05) was taken for age 1 and age 2 of that time series only consisted of landings.

Efforts were made at WGNSSK 2016 to revise the stock weights from 2012 but because of unrealistic submitted sample data from some countries this exercise was unsuccessful and only 2015 was corrected.

### 10.2.6 Maturity and natural mortality

A knife-edged maturity-ogive with full maturation at age 3 is assumed for North Sea sole (Table 10.2.7.).No new data was presented at WGNSKK 2016.
Natural mortality at age (Table 10.2.7.) has been assumed to be constant at 0.1, except for 1963 where a value of 0.9 was used to take into account the effect of the severe winter (1962-1963) (ICES-FWG 1979). The estimate of 0.9 was based on an analysis of CPUE in the fisheries before and after the severe winter (CM 1979/G:10).

### 10.2.7 Catch, effort and survey data

Two tuning series that take place in quarter 3 are used in the assessment. The BTS-ISIS (Beam Trawl Survey on the RV ISIS) and the SNS (Sole Net Survey) are both surveys conducted by the Netherlands. Catches of sole in the 2012 survey were extremely low and contradicted with the BTS, indicating problems with operating the gear properly on board of the vessel. The data from the SNS survey for the years 2003 and 2012 were not made available.

The BTS-ISIS and SNS 2015 surveys show large yearclasses of 2009-2011 coming through in ages 6-9.

A standardised comparison of the two surveys that are used as tuning indices over the available time series is given in Figure 10.2.7.1. The internal consistency of the year class cohorts in these two surveys is presented in Figure 10.2.7.2.

An additional survey index (the combined Belgian, German, and Dutch DFS0) is used for recruitment estimates in the RCT3 analysis.

All survey indices of importance for the advice are presented in Table 10.2.8.
In autumn, when new data becomes available from the surveys in quarter 3, the advice can be revised if significant changes in the assumptions of recruitment made at WGNSSK 2016 are observed.

### 10.3 Assessment

The model used is the Art and Poos model (AAP, Aarts and Poos (2009), for more details please refer to the Stock Annex).

|  | Year of ASSESSMENT: |
| :--- | :--- |
| Assessment model: | AAP |
| Assessment software | FLR/ADMB |
| Fleets: |  |
| BTS-ISIS Age range |  |
| $\quad$ Year range | $1-9$ |
| Age range <br> Year range | $1985-$ present |
| SNS | $1-6$ |
| Catch/Landings | $1970-$ present |
| Age range: |  |
| Landings data: | $1-10+$ |
| Discards data | $1957-$ present |
| Model settings | 2002 -present |
| Fbar: |  |
| Age from which F is constant (qplat.Fmatrix) | $2-6$ |
| Dimension of the F matrix (Fage.knots) | 8 |
| Ftime.knots | 6 |
| Wtime.knots | 22 |
| Age from which q is constant (qplat.surveys) | 5 |

This is an update assessment with, in principle, only an update of historical data and addition of the commercial and survey data in the most recent year. The model settings, defined in the most recent benchmark by WKNSEA (2015), were applied.

The assessment summary is presented in Table 10.3.1. and in Figure 10.3.1. The retrospective performance of the assessment is shown in Figure 10.3.2.

### 10.4 Recruitment estimates

Recruitment estimation was carried out using RCT3. Input to the RCT3 model is presented in Table 10.4.1. Results are presented in Table 10.4.2. for age 1 and Table 10.4.3. for age 2. Average recruitment of 1-year old fish in the period 1957-2012 was around 111 million (geometric mean).

The results are summarized in the table below and the estimates used for the shortterm forecast are underlined.

|  |  | AAP |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Year Class | AGe in 2016 | Thousands | RCT3 <br> Thousands | GM(1957-2012) <br> Thousands |
| 2014 | 2 | 163431 | 137233 | 99023 |
| 2015 | 1 |  | 59248 | 111851 |
| 2016 | Recruit $(0)$ |  |  | 111851 |

Additional recruitment information will be available from the 3rd quarter surveys (BTS-ISIS, SNS, and DFS) carried out in 2016. ICES will only issue an updated advice if these surveys provide a very different perspective on the short-term developments.

### 10.5 Short-term forecasts

The short-term forecasts were carried out with FLR. The exploitation pattern (F) was taken to be the mean value of the last three years. Weight-at-age in the stock and weight-at-age in the catch were taken to be the mean of the last three years. Population numbers at ages 2 and older are AAP survivor estimates. Numbers at age 1 are taken from the RCT3 analysis and recruitment of the 2015 and later year-classes are taken from the long-term geometric mean (1957-2011: 94 million). Input to and results from the short term forecast are presented in Table 10.5.1.-3. for $\mathrm{F}=\mathrm{Fsq}$ and Table 10.5.4-6. for catch $=$ TAC.

For the intermediate year 2016, it was assumed that catches equal the TAC at WGNSSK 2016 (since North Sea sole is under the landings obligation in 2016). The expected landings in 2015 of 12761 t are close to the agreed TAC of 2015 (12 262 t ). Therefore the landings in 2016 are assuming that the TAC will be fully taken. This corresponds with the observations in recent years.

Figure 10.5.1-2. shows the relative contribution of assumptions under both scenario's.

### 10.6 Medium-term forecasts

No medium term projections were done this year.

### 10.7 Biological reference points

|  | Type | Value | Technical basis |
| :---: | :---: | :---: | :---: |
| MSY <br> Approach | MSY <br> Btrigger | 37000 t | Default to value of Bpa |
|  | FMSY | 0.2 | Median of stochastic MSY analysis assuming a Hockeystick stock-recruit relationship. |
|  | Fmsyupper | 0.37 |  |
|  | Fmsylower | 0.11 |  |
| Precautionary Approach | Blim | 26300 t | Breakpoint of segmented regression (WKNSEA 2015) |
|  | Bpa | 37000 t | Bpa $=1.4$ * Blim |
|  | Flim | 0.63 | EqSim run with no MSY Btrigger, realistic assessment.advice error, biological parameters (20032012) and fishery parameters (2009-2012) |
|  | Fpa | 0.44 | Fpa $=$ Flim / 1.4 |

## FMSY reference points

In 2010 ICES implemented the MSY framework for providing advice on the exploitation of stocks. The aim is to manage all stocks at an exploitation rate ( F ) that is consistent with maximum (high) long term yield while providing a low risk to the stock.
In 2014 the joint ICES-MYFISH Workshop (WKMSYREF3, ICES 2014) held place to consider the basis for FMSY ranges of, among others, SOL4. The workshop convened again under the auspices of WKLIFE in March 2015. This eventually resulted in an $\mathrm{F}_{\text {msy }}$ range for sole of 0.13-0.27. The point value of $\mathrm{F}_{\text {msy }}$ was set at 0.2 .

At WGNSSK 2016, $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{F}_{\text {lim }}$ were defined according to ICES reference points guidelines $(\mathrm{ACOM})$. An additional $\mathrm{F}_{\mathrm{pa}}($ sigma $)$ was estimated by: $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\lim } / \exp (1.645$ *
sigma), where sigma is the standard deviation of $\ln (\mathrm{F})$ in the final assessment year. $\mathrm{F}_{\mathrm{pa}}$ (sigma) was estimated as 0.48 .

### 10.8 Quality of the assessment

The assessment was benchmarked recently in February 2015 (WKNSEA, 2015). Inclusion of discards in the catches and adding uncertainty estimates were the main goals. This was attained using the AAP-model.
Discards form a minor part of total sole catches, rates have stabilised in the last years. The assessment at present includes 15 years of discards data obtained from sampling programs in several countries and is considered to be robust and consistent between years.

Most of the discards originate from the Netherlands. A self-sampling programme by the Dutch beam-trawl fleet has been in place since 2004. This sampling programme indicates spatial and temporal trends in discarding (higher discards are observed in coastal regions and late summer), but it was considered inappropriate for overall estimates of discarding because of differences in the implementations of sampling methods.

In 2009, a new self-sampling programme was launched to address this. Since 2011, Dutch discard estimates are derived exclusively from the self-sampling programme, while observer estimates are used for validation of the self-sampling data only. Preliminary analyses suggest that the self-sampling estimates are as reliable as those from the observer programme (Chen et al. (in press)).

At WGNSSK 2016 the newest data year (2015) was added to the assessment. The assessment performed well and modelled landings, discards, and catch fitted well to observed landings, and discards (Figure 10.8.1-3.). It is apparent that the AAP-model estimates of the landings are slightly overestimated in the recent years. Whereas discards are slightly underestimated.

Residual plots of landings and discards are shown in Figure 10.8.4.-5. Residuals are small for younger ages in discards but tend to be higher for older ages. This is normal since older North Sea sole are not seen in discards.

Sigmas of the different data time series are shown in Figure 10.8.6.

### 10.9 Status of the Stock

Fishing mortality was estimated at 0.20 in 2015 which is well within biological limits and on F msy $(0.2)$. The SSB in 2015 was estimated at about 49142 t which is well above both $\mathrm{B}_{\lim }$ and $\mathrm{B}_{\mathrm{pa}}$.

### 10.10 Management Considerations

Sole is mainly taken by beam trawlers in a mixed fishery for sole and plaice in the southern and central part of the North Sea. The long term management plan for plaice and sole in the North Sea specifies two distinct phases. The objective of stage one of the flatfish management plan was to bring both sole and plaice stocks within safe biological limits. This objective has been achieved for both stocks. The plaice stock is estimated above $B_{p a}$ since 2005 and the sole stock is above $B_{p a}$ since 2012. Also fishing mortalities are well below $\mathrm{F}_{\mathrm{pa}}$ for both stocks for a number of years.

The management plan foresees a re-evaluation of the biological objectives and introduction of economic and social objectives after stage 1 is completed. The management
plan states that when the stocks of plaice and sole have been found for two years in succession to have returned to within safe biological limits, the Council shall decide on the basis of a proposal from the Commission on the amendment of Articles 4(2) and 4(3) and the amendment of Articles 7, 8 and 9 that will, in the light of the latest scientific advice from the STECF, permit the exploitation of the stocks at a fishing mortality rate compatible with maximum sustainable yield.

The management plan is in stage 2 now and action should be taken to specify the implementation in this stage. The multiannual plan states that, in its second stage, it shall ensure the exploitation of the stocks of plaice and sole on the basis of maximum sustainable yield. An overall objective of the CFP is to aim exploitation of all fish stocks at $\mathrm{F}_{\text {msy }}$.

The majority of the sole catches are taken by beam trawlers in a mixed fishery with other flatfish and roundfish species. In general discards of other species in beam trawls are rather high. Due to measures resulting from the flatfish management plan, actions taken to reduce bycatch, disturbance to the sea bottom, and economic incentives (reduce fuel costs), overall effort in the beam fishery has been reduced in the past 16 years by $70 \%$. The significant reduction of effort in the fleet must have contributed to reduce the impact of this fishery on the marine ecosystem.

### 10.11 Frequency of assessment

The frequency of assessments was discussed at the ACOM December 2014 meeting and the Committee decided to develop simple criteria to be used to identify stocks that would be candidates for less frequent assessments. A set of four criteria were suggested based on (1) the life span of the stock, (2) stock status, (3) relative importance of recruitment in the catch forecast and (4) the quality of the assessment.

At WGNSSK 2015 the four criteria were assessed. The North Sea sole assessment succeeded in all four criteria. Although the North Sea sole stock is consequently a candidate for less frequent assessments some precautions should be taken in to account:

- North Sea sole is subject to the landing obligation as of 2016, this implies careful proceeding with discard data that are input for the model.
- Furthermore, the main fleet targeting sole is subject to technological changes in their gears. How this technological change affects the selectivity of the fishing gears catching sole and subsequently the age composition of the stock has not been quantified.
- Finally, the assessment currently holds two tuning indices that are not encompassing the whole sole stock in the North Sea and are missing out on the main grounds where sole is found. The positive trend in the assessment and its basis thereof for the second criterion on the frequency of assessment should be therefore taken with caution.


## Criterion

North Sea sole
(1) Life span (i.e. maximum normal age) of the species Life span larger than 5 years is larger than 5 years
(2) The stock status in relation to the reference points is $\mathrm{F}(2015)=0.20<\mathrm{F}_{\text {upper }}$ according to the MSY criteria F (latest assessment year) is $<=\mathrm{F}_{\text {upper }}$ (upper bound in F range) AND SSB (start of intermediate year) $>=$ MSY $B_{\text {trigger }}$
(3) The average contribution to the catch in numbers of the recruiting year class in latest 5 years is less than $25 \%$ of the total catch in numbers.
(4) The retrospective pattern, based on a seven years peel of Mohn's Rho index, shows that F is consistently overestimated by less than $20 \%$

The average contribution to the catch in numbers of the recruiting year class in latest 5 years is $19 \%$ of the total catch in numbers

$$
\text { Rho }=-0.1
$$

i.e. F is overestimated by $10 \%$

Table 10.2.1 North Sea sole. Landings per country, total reported landings, ICES total landings, and TAC

| Year | BE | DK | FR | GE | NL | UK | Other | Total REPORTED LANDINGS | ICES <br> Total <br> LANDINGS | TAC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 1900 | 524 | 686 | 266 | 17686 | 403 | 2 | 21467 | 21579 | 21000 |
| 1983 | 1740 | 730 | 332 | 619 | 16101 | 435 |  | 19957 | 24927 | 20000 |
| 1984 | 1771 | 818 | 400 | 1034 | 14330 | 586 | 1 | 18940 | 26839 | 20000 |
| 1985 | 2390 | 692 | 875 | 303 | 14897 | 774 | 3 | 19934 | 24248 | 22000 |
| 1986 | 1833 | 443 | 296 | 155 | 9558 | 647 | 2 | 12934 | 18201 | 20000 |
| 1987 | 1644 | 342 | 318 | 210 | 10635 | 676 | 4 | 13829 | 17368 | 14000 |
| 1988 | 1199 | 616 | 487 | 452 | 9841 | 740 | 28 | 13363 | 21590 | 14000 |
| 1989 | 1596 | 1020 | 312 | 864 | 9620 | 1033 | 50 | 14495 | 21805 | 14000 |
| 1990 | 2389 | 1427 | 352 | 2296 | 18202 | 1614 | 263 | 26543 | 35120 | 25000 |
| 1991 | 2977 | 1307 | 465 | 2107 | 18758 | 1723 | 271 | 27608 | 33513 | 27000 |
| 1992 | 2058 | 1359 | 548 | 1880 | 18601 | 1281 | 277 | 26004 | 29341 | 25000 |
| 1993 | 2783 | 1661 | 490 | 1379 | 22015 | 1149 | 298 | 29775 | 31491 | 32000 |
| 1994 | 2935 | 1804 | 499 | 1744 | 22874 | 1137 | 298 | 31291 | 33002 | 32000 |
| 1995 | 2624 | 1673 | 640 | 1564 | 20927 | 1040 | 312 | 28780 | 30467 | 28000 |
| 1996 | 2555 | 1018 | 535 | 670 | 15344 | 848 | 229 | 21199 | 22651 | 23000 |
| 1997 | 1519 | 689 | 99 | 510 | 10241 | 479 | 204 | 13741 | 14901 | 18000 |
| 1998 | 1844 | 520 | 510 | 782 | 15198 | 549 | 339 | 19742 | 20868 | 19100 |
| 1999 | 1919 | 828 | NA | 1458 | 16283 | 645 | 501 | 21634 | 23475 | 22000 |
| 2000 | 1806 | 1069 | 362 | 1280 | 15273 | 600 | 539 | 20929 | 22641 | 22000 |
| 2001 | 1874 | 772 | 411 | 958 | 13345 | 597 | 394 | 18351 | 19944 | 19000 |
| 2002 | 1437 | 644 | 266 | 759 | 12120 | 451 | 292 | 15969 | 16945 | 16000 |
| 2003 | 1605 | 703 | 728 | 749 | 12469 | 521 | 363 | 17138 | 17920 | 15850 |
| 2004 | 1477 | 808 | 655 | 949 | 12860 | 535 | 544 | 17828 | 18757 | 17000 |
| 2005 | 1374 | 831 | 676 | 756 | 10917 | 667 | 357 | 15579 | 16355 | 18600 |
| 2006 | 980 | 585 | 648 | 475 | 8299 | 910 | 0 | 11933 | 12594 | 17670 |
| 2007 | 955 | 413 | 401 | 458 | 10365 | 1203 | 5 | 13800 | 14635 | 15000 |
| 2008 | 1379 | 507 | 714 | 513 | 9456 | 851 | 15 | 13435 | 14071 | 12800 |
| 2009 | 1353 | NA | NA | 555 | 12038 | 951 | 1 | NA | 13952 | 14000 |
| 2010 | 1268 | 406 | 621 | 537 | 8770 | 526 | 1.38 | 12129 | 12603 | 14100 |
| 2011 | 857 | 346 | 539 | 327 | 8133 | 786 | 2 | 10990 | 11485 | 14100 |
| 2012 | 593 | 418 | 633 | 416 | 9089 | 599 | 3 | 11752 | 11602 | 16200 |
| 2013 | 697 | 497 | 680 | 561 | 9987 | 867 | 0 | 13291 | 13137 | 14000 |
| 2014 | 920 | 314 | 675 | 642 | 9569 | 840 | 0 | 12547 | 13060 | 11900 |
| 2015 | 933 | 271 | 532 | 765 | 8899 | 804 | 0 | 12203 | 12867 | 11900 |

Table 10.2.2. North Sea sole. Landings in numbers by age (in thousands) as input for the assessment model, age 10 is a plusgroup

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1957 | 0 | 1472 | 10556 | 13150 | 3913 | 3041 | 6780 | 1803 | 529 | 6541 |
| 1958 | 0 | 1863 | 8482 | 14240 | 9547 | 3501 | 3023 | 4461 | 2264 | 6590 |
| 1959 | 0 | 3694 | 12139 | 10499 | 9060 | 5823 | 1217 | 2044 | 2598 | 5668 |
| 1960 | 0 | 11965 | 14043 | 16691 | 9248 | 8313 | 4815 | 1583 | 1049 | 7851 |
| 1961 | 0 | 972 | 50470 | 19403 | 12574 | 4760 | 3998 | 4338 | 847 | 7355 |
| 1962 | 0 | 1584 | 6173 | 58836 | 15254 | 10478 | 4797 | 4087 | 2074 | 7450 |
| 1963 | 0 | 670 | 8271 | 8485 | 45823 | 8420 | 6603 | 2403 | 3365 | 8316 |
| 1964 | 53 | 150 | 2041 | 5518 | 3680 | 16749 | 3020 | 1749 | 790 | 2913 |
| 1965 | 0 | 45180 | 1045 | 1534 | 4798 | 2381 | 11990 | 1494 | 1463 | 3077 |
| 1966 | 0 | 12145 | 132170 | 979 | 1168 | 3649 | 736 | 6255 | 694 | 2424 |
| 1967 | 0 | 3769 | 26260 | 87039 | 1998 | 548 | 1962 | 777 | 5160 | 2978 |
| 1968 | 1034 | 17093 | 13852 | 24894 | 48417 | 461 | 244 | 1639 | 323 | 6502 |
| 1969 | 404 | 24404 | 21884 | 5433 | 12638 | 25646 | 338 | 249 | 1214 | 5379 |
| 1970 | 1299 | 6141 | 25996 | 8236 | 1784 | 3231 | 11961 | 246 | 140 | 5234 |
| 1971 | 425 | 33765 | 14596 | 12909 | 4538 | 1459 | 2355 | 7300 | 194 | 4649 |
| 1972 | 354 | 7511 | 36356 | 6997 | 4911 | 1548 | 517 | 1218 | 4654 | 2772 |
| 1973 | 716 | 12459 | 13025 | 16493 | 4101 | 2368 | 1013 | 779 | 1241 | 5899 |
| 1974 | 100 | 15171 | 21248 | 5412 | 6965 | 1896 | 1563 | 649 | 396 | 4750 |
| 1975 | 267 | 23193 | 28833 | 11839 | 2110 | 3870 | 798 | 916 | 513 | 3481 |
| 1976 | 1064 | 3619 | 28571 | 14316 | 4923 | 987 | 1950 | 562 | 434 | 2721 |
| 1977 | 1780 | 22747 | 12299 | 15593 | 7580 | 1812 | 325 | 1133 | 261 | 2155 |
| 1978 | 27 | 24921 | 29163 | 6102 | 6610 | 4231 | 1730 | 608 | 643 | 1595 |
| 1979 | 9 | 8280 | 41681 | 16259 | 3033 | 3262 | 1769 | 826 | 244 | 1546 |
| 1980 | 650 | 1233 | 12762 | 18138 | 7444 | 1479 | 2241 | 1437 | 374 | 1227 |
| 1981 | 434 | 29983 | 3344 | 7046 | 8439 | 3757 | 973 | 909 | 786 | 932 |
| 1982 | 2697 | 26799 | 46375 | 1868 | 3584 | 4855 | 1701 | 623 | 613 | 1295 |
| 1983 | 391 | 34545 | 41551 | 21273 | 626 | 1383 | 1958 | 982 | 388 | 1181 |
| 1984 | 192 | 30839 | 44081 | 22631 | 8821 | 744 | 857 | 1047 | 526 | 897 |
| 1985 | 163 | 16449 | 42773 | 20079 | 9307 | 3520 | 207 | 375 | 631 | 965 |
| 1986 | 372 | 9304 | 18381 | 17591 | 7698 | 5480 | 2256 | 109 | 281 | 1671 |
| 1987 | 93 | 28896 | 21927 | 8851 | 6477 | 3102 | 1559 | 898 | 81 | 690 |
| 1988 | 10 | 13206 | 47135 | 15217 | 4377 | 3878 | 1549 | 890 | 523 | 317 |
| 1989 | 115 | 45652 | 17973 | 22295 | 4551 | 1627 | 1414 | 637 | 451 | 459 |
| 1990 | 854 | 11816 | 103380 | 9667 | 9099 | 3315 | 1032 | 1186 | 548 | 837 |
| 1991 | 118 | 12938 | 24985 | 76580 | 6609 | 3612 | 1706 | 707 | 718 | 1072 |
| 1992 | 965 | 6730 | 43713 | 15961 | 37745 | 2440 | 2995 | 730 | 393 | 1163 |
| 1993 | 53 | 49870 | 16575 | 31047 | 13709 | 23758 | 1472 | 1170 | 456 | 833 |
| 1994 | 709 | 7710 | 86349 | 13387 | 18513 | 5642 | 11174 | 458 | 905 | 897 |
| 1995 | 4766 | 12674 | 16700 | 68073 | 6262 | 7254 | 1981 | 5971 | 293 | 665 |
| 1996 | 170 | 18609 | 16005 | 16770 | 26946 | 3814 | 4725 | 932 | 3267 | 976 |
| 1997 | 1574 | 5987 | 23418 | 7253 | 5058 | 12667 | 1189 | 2303 | 330 | 1672 |
| 1998 | 242 | 56162 | 15011 | 14806 | 3466 | 1924 | 4727 | 787 | 1022 | 838 |


|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 284 | 15601 | 71730 | 8103 | 6049 | 1200 | 657 | 1964 | 328 | 804 |
| 2000 | 2329 | 14929 | 32425 | 42394 | 3257 | 2453 | 796 | 431 | 922 | 708 |
| 2001 | 857 | 25045 | 20925 | 19260 | 16211 | 1383 | 808 | 266 | 163 | 701 |
| 2002 | 1046 | 10958 | 32570 | 12185 | 8145 | 6393 | 667 | 592 | 88 | 362 |
| 2003 | 1047 | 32295 | 17479 | 16072 | 5814 | 3902 | 2427 | 400 | 128 | 451 |
| 2004 | 516 | 14960 | 48003 | 9531 | 7462 | 2167 | 902 | 962 | 389 | 389 |
| 2005 | 1131 | 7254 | 22633 | 28875 | 4168 | 3861 | 1491 | 602 | 768 | 392 |
| 2006 | 7008 | 9966 | 10397 | 9606 | 10943 | 1617 | 1577 | 724 | 373 | 553 |
| 2007 | 315 | 39643 | 10820 | 6407 | 5706 | 5479 | 819 | 725 | 498 | 541 |
| 2008 | 1959 | 6325 | 37427 | 5996 | 2928 | 2393 | 2613 | 448 | 491 | 459 |
| 2009 | 1630 | 10417 | 10771 | 26548 | 3278 | 1652 | 1591 | 1532 | 312 | 864 |
| 2010 | 371 | 11659 | 13354 | 8530 | 13623 | 1817 | 907 | 809 | 1196 | 690 |
| 2011 | 44 | 11992 | 19788 | 8379 | 5070 | 6436 | 983 | 431 | 283 | 765 |
| 2012 | 1 | 6439 | 28605 | 11069 | 4285 | 2146 | 4072 | 587 | 286 | 1028 |
| 2013 | 0 | 2741 | 28189 | 21500 | 5643 | 2042 | 1532 | 2246 | 242 | 471 |
| 2014 | 371 | 8111 | 6916 | 22942 | 11440 | 2591 | 1808 | 620 | 840 | 459 |
| 2015 | 201 | 10512 | 16589 | 4738 | 14756 | 6157 | 1470 | 562 | 393 | 545 |

Table 10.2.3. North Sea sole. Discards in numbers by age (in thousands) as input for the assessment model, age 10 is a plusgroup

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 6461 | 12606 | 5212 | 1029 | 272 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 1156 | 7152 | 5059 | 1212 | 381 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 2936 | 12832 | 7449 | 1719 | 518 | 12 | 0 | 0 | 0 | 0 |
| 2005 | 2256 | 5622 | 4796 | 1258 | 375 | 63 | 22 | 0 | 0 | 0 |
| 2006 | 2390 | 5727 | 2705 | 654 | 197 | 28 | 18 | 7 | 0 | 0 |
| 2007 | 818 | 4923 | 3010 | 619 | 226 | 57 | 4 | 0 | 0 | 0 |
| 2008 | 1230 | 2704 | 1764 | 371 | 106 | 0 | 8 | 0 | 0 | 0 |
| 2009 | 2695 | 6480 | 3652 | 999 | 266 | 5 | 9 | 0 | 0 | 0 |
| 2010 | 5687 | 12164 | 6670 | 1544 | 493 | 31 | 10 | 2 | 2 | 0 |
| 2011 | 3457 | 10298 | 5482 | 1273 | 354 | 33 | 0 | 0 | 0 | 0 |
| 2012 | 1132 | 19556 | 9444 | 984 | 230 | 232 | 36 | 4 | 7 | 1 |
| 2013 | 4653 | 5733 | 12558 | 3649 | 340 | 125 | 19 | 3 | 0 | 0 |
| 2014 | 7162 | 5836 | 2371 | 3488 | 1366 | 238 | 198 | 6 | 0 | 0 |
| 2015 | 9454 | 9166 | 3913 | 1991 | 1528 | 415 | 15 | 50 | 8 | 1 |

Table 10.2.4. North Sea sole. Landings weights (kg) at age as input for the assessment model, age 10 is a plusgroup

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1957 | 0.155 | 0.154 | 0.177 | 0.204 | 0.248 | 0.279 | 0.29 | 0.335 | 0.436 | 0.40813 |
| 1958 | 0.155 | 0.145 | 0.178 | 0.22 | 0.254 | 0.273 | 0.314 | 0.323 | 0.388 | 0.41344 |
| 1959 | 0.155 | 0.162 | 0.188 | 0.228 | 0.261 | 0.301 | 0.328 | 0.321 | 0.373 | 0.42621 |
| 1960 | 0.155 | 0.153 | 0.185 | 0.235 | 0.254 | 0.277 | 0.301 | 0.309 | 0.381 | 0.4177 |
| 1961 | 0.155 | 0.146 | 0.174 | 0.211 | 0.255 | 0.288 | 0.319 | 0.304 | 0.346 | 0.41932 |
| 1962 | 0.155 | 0.155 | 0.165 | 0.208 | 0.241 | 0.295 | 0.32 | 0.321 | 0.334 | 0.41186 |
| 1963 | 0.155 | 0.163 | 0.171 | 0.219 | 0.258 | 0.309 | 0.323 | 0.387 | 0.376 | 0.48463 |
| 1964 | 0.153 | 0.175 | 0.213 | 0.252 | 0.274 | 0.309 | 0.327 | 0.346 | 0.388 | 0.4805 |
| 1965 | 0.155 | 0.169 | 0.209 | 0.246 | 0.286 | 0.282 | 0.345 | 0.378 | 0.404 | 0.47972 |
| 1966 | 0.155 | 0.177 | 0.19 | 0.18 | 0.301 | 0.332 | 0.429 | 0.399 | 0.449 | 0.50148 |
| 1967 | 0.155 | 0.192 | 0.201 | 0.252 | 0.277 | 0.389 | 0.419 | 0.339 | 0.424 | 0.49123 |
| 1968 | 0.157 | 0.189 | 0.207 | 0.267 | 0.327 | 0.342 | 0.354 | 0.455 | 0.465 | 0.50752 |
| 1969 | 0.152 | 0.191 | 0.196 | 0.255 | 0.311 | 0.373 | 0.553 | 0.398 | 0.468 | 0.52271 |
| 1970 | 0.154 | 0.212 | 0.218 | 0.285 | 0.35 | 0.404 | 0.441 | 0.463 | 0.443 | 0.5326 |
| 1971 | 0.145 | 0.193 | 0.237 | 0.322 | 0.358 | 0.425 | 0.42 | 0.49 | 0.534 | 0.54714 |
| 1972 | 0.169 | 0.204 | 0.252 | 0.334 | 0.434 | 0.425 | 0.532 | 0.485 | 0.558 | 0.62907 |
| 1973 | 0.146 | 0.208 | 0.238 | 0.346 | 0.404 | 0.448 | 0.552 | 0.567 | 0.509 | 0.58575 |
| 1974 | 0.164 | 0.192 | 0.233 | 0.338 | 0.418 | 0.448 | 0.52 | 0.559 | 0.609 | 0.65327 |
| 1975 | 0.129 | 0.182 | 0.225 | 0.32 | 0.406 | 0.456 | 0.529 | 0.595 | 0.629 | 0.66935 |
| 1976 | 0.143 | 0.19 | 0.222 | 0.306 | 0.389 | 0.441 | 0.512 | 0.562 | 0.667 | 0.66472 |
| 1977 | 0.147 | 0.188 | 0.236 | 0.307 | 0.369 | 0.424 | 0.43 | 0.52 | 0.562 | 0.6194 |
| 1978 | 0.152 | 0.196 | 0.231 | 0.314 | 0.37 | 0.426 | 0.466 | 0.417 | 0.572 | 0.66635 |
| 1979 | 0.137 | 0.208 | 0.246 | 0.323 | 0.391 | 0.448 | 0.534 | 0.544 | 0.609 | 0.76296 |
| 1980 | 0.141 | 0.199 | 0.244 | 0.331 | 0.371 | 0.418 | 0.499 | 0.55 | 0.598 | 0.68412 |
| 1981 | 0.143 | 0.187 | 0.226 | 0.324 | 0.378 | 0.424 | 0.442 | 0.516 | 0.542 | 0.63022 |
| 1982 | 0.141 | 0.188 | 0.216 | 0.307 | 0.371 | 0.409 | 0.437 | 0.491 | 0.58 | 0.65568 |
| 1983 | 0.134 | 0.182 | 0.217 | 0.301 | 0.389 | 0.416 | 0.467 | 0.489 | 0.505 | 0.64225 |
| 1984 | 0.153 | 0.171 | 0.221 | 0.286 | 0.361 | 0.386 | 0.465 | 0.555 | 0.575 | 0.63382 |
| 1985 | 0.122 | 0.187 | 0.216 | 0.288 | 0.357 | 0.427 | 0.447 | 0.544 | 0.612 | 0.64476 |
| 1986 | 0.135 | 0.179 | 0.213 | 0.299 | 0.357 | 0.407 | 0.485 | 0.543 | 0.568 | 0.60955 |
| 1987 | 0.139 | 0.185 | 0.205 | 0.277 | 0.356 | 0.378 | 0.428 | 0.481 | 0.393 | 0.65696 |
| 1988 | 0.127 | 0.175 | 0.217 | 0.27 | 0.354 | 0.428 | 0.484 | 0.521 | 0.559 | 0.71241 |
| 1989 | 0.118 | 0.173 | 0.216 | 0.288 | 0.336 | 0.375 | 0.456 | 0.492 | 0.47 | 0.61107 |
| 1990 | 0.124 | 0.183 | 0.227 | 0.292 | 0.371 | 0.413 | 0.415 | 0.514 | 0.476 | 0.61975 |
| 1991 | 0.127 | 0.186 | 0.21 | 0.263 | 0.315 | 0.436 | 0.443 | 0.467 | 0.507 | 0.55809 |
| 1992 | 0.146 | 0.178 | 0.213 | 0.258 | 0.298 | 0.38 | 0.409 | 0.46 | 0.487 | 0.55569 |
| 1993 | 0.097 | 0.167 | 0.196 | 0.239 | 0.264 | 0.3 | 0.338 | 0.441 | 0.496 | 0.60312 |
| 1994 | 0.143 | 0.18 | 0.202 | 0.228 | 0.257 | 0.3 | 0.317 | 0.432 | 0.409 | 0.51009 |
| 1995 | 0.151 | 0.186 | 0.196 | 0.247 | 0.265 | 0.319 | 0.344 | 0.356 | 0.444 | 0.59158 |
| 1996 | 0.163 | 0.177 | 0.202 | 0.234 | 0.274 | 0.285 | 0.318 | 0.37 | 0.39 | 0.59428 |
| 1997 | 0.151 | 0.18 | 0.206 | 0.236 | 0.267 | 0.296 | 0.323 | 0.306 | 0.384 | 0.4396 |
| 1998 | 0.128 | 0.182 | 0.189 | 0.252 | 0.262 | 0.289 | 0.336 | 0.292 | 0.335 | 0.50367 |


|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 0.163 | 0.179 | 0.212 | 0.229 | 0.287 | 0.324 | 0.354 | 0.372 | 0.372 | 0.45268 |
| 2000 | 0.145 | 0.17 | 0.2 | 0.248 | 0.29 | 0.299 | 0.323 | 0.368 | 0.402 | 0.42761 |
| 2001 | 0.143 | 0.185 | 0.202 | 0.27 | 0.275 | 0.333 | 0.391 | 0.414 | 0.433 | 0.49344 |
| 2002 | 0.14 | 0.183 | 0.211 | 0.243 | 0.281 | 0.312 | 0.366 | 0.319 | 0.571 | 0.53635 |
| 2003 | 0.136 | 0.182 | 0.214 | 0.256 | 0.273 | 0.317 | 0.34 | 0.344 | 0.503 | 0.43054 |
| 2004 | 0.127 | 0.18 | 0.209 | 0.252 | 0.263 | 0.284 | 0.378 | 0.367 | 0.327 | 0.42456 |
| 2005 | 0.172 | 0.185 | 0.207 | 0.243 | 0.241 | 0.282 | 0.265 | 0.377 | 0.318 | 0.40057 |
| 2006 | 0.156 | 0.19 | 0.22 | 0.263 | 0.291 | 0.322 | 0.293 | 0.358 | 0.397 | 0.39622 |
| 2007 | 0.154 | 0.18 | 0.205 | 0.237 | 0.253 | 0.273 | 0.295 | 0.299 | 0.281 | 0.32644 |
| 2008 | 0.15 | 0.181 | 0.223 | 0.24 | 0.265 | 0.324 | 0.314 | 0.297 | 0.307 | 0.41748 |
| 2009 | 0.138 | 0.185 | 0.202 | 0.256 | 0.275 | 0.278 | 0.325 | 0.334 | 0.303 | 0.39787 |
| 2010 | 0.163 | 0.181 | 0.22 | 0.236 | 0.273 | 0.308 | 0.283 | 0.311 | 0.361 | 0.38068 |
| 2011 | 0.152 | 0.162 | 0.194 | 0.233 | 0.242 | 0.274 | 0.272 | 0.293 | 0.335 | 0.34695 |
| 2012 | 0.095 | 0.169 | 0.185 | 0.233 | 0.256 | 0.234 | 0.27 | 0.26 | 0.283 | 0.269 |
| 2013 | 0.125 | 0.169 | 0.185 | 0.224 | 0.253 | 0.266 | 0.297 | 0.278 | 0.309 | 0.466 |
| 2014 | 0.155 | 0.191 | 0.212 | 0.228 | 0.263 | 0.273 | 0.249 | 0.279 | 0.319 | 0.351 |
| 2015 | 0.145 | 0.169 | 0.205 | 0.24 | 0.263 | 0.274 | 0.304 | 0.293 | 0.33 | 0.31934 |

Table 10.2.5. North Sea sole. Discard weights (kg) at age as input for the assessment model, age 10 is a plusgroup

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 0.046 | 0.068 | 0.084 | 0.091 | 0.096 | 0.11 | 0.124 | 0.137 | 0.137 | 0 |
| 2003 | 0.054 | 0.087 | 0.1 | 0.107 | 0.114 | 0.11 | 0.124 | 0.137 | 0.137 | 0 |
| 2004 | 0.065 | 0.089 | 0.103 | 0.111 | 0.118 | 0.095 | 0.124 | 0.137 | 0.137 | 0 |
| 2005 | 0.068 | 0.089 | 0.104 | 0.109 | 0.114 | 0.103 | 0.107 | 0.137 | 0.137 | 0 |
| 2006 | 0.066 | 0.082 | 0.099 | 0.109 | 0.108 | 0.115 | 0.113 | 0.121 | 0.137 | 0 |
| 2007 | 0.066 | $0.087$ | 0.098 | 0.102 | 0.107 | 0.104 | 0.121 | 0.136 | 0.136 | 0 |
| 2008 | 0.064 | 0.086 | 0.101 | 0.112 | 0.124 | 0.11 | 0.111 | 0.137 | 0.137 | 0 |
| $2009$ | 0.066 | 0.089 | 0.101 | 0.106 | 0.114 | 0.126 | 0.104 | 0.137 | 0.137 | 0 |
| 2010 | 0.066 | 0.083 | 0.096 | 0.105 | 0.109 | 0.111 | 0.113 | 0.121 | 0.121 | 0 |
| 2011 | 0.053 | 0.081 | 0.093 | 0.104 | 0.113 | 0.104 | 0.11 | 0.122 | 0.126 | 0 |
| 2012 | 0.059 | 0.075 | 0.09 | 0.096 | 0.111 | 0.08 | 0.115 | 0.122 | 0.121 | 0.14 |
| 2013 | 0.041 | 0.075 | 0.086 | 0.1 | 0.117 | 0.09 | 0.112 | 0.117 | 0.121 | 0 |
| 2014 | 0.051 | 0.079 | 0.089 | 0.097 | 0.106 | 0.1 | 0.117 | 0.099 | 0.147 | 0 |
| 2015 | 0.032 | 0.076 | 0.095 | 0.087 | 0.105 | 0.117 | 0.132 | 0.124 | 0.159 | 0.199 |

Table 10.2.6. North Sea sole. Stock weights ( $\mathbf{k g}$ ) at age $(\mathbf{k g})$ as input for the assessment model, age 10 is a plusgroup. Mean weights of sampled catches in quarter 2 are exported from InterCatch. Danish weights were not extracted for quarter 2 at WGNSSK 2016

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1957 | 0.025 | 0.07 | 0.147 | 0.187 | 0.208 | 0.253 | 0.262 | 0.355 | 0.39 | 0.36517 |
| 1958 | 0.025 | 0.07 | 0.164 | 0.205 | 0.226 | 0.228 | 0.297 | 0.318 | 0.393 | 0.4215 |
| 1959 | 0.025 | 0.07 | 0.159 | 0.198 | 0.239 | 0.271 | 0.292 | 0.276 | 0.303 | 0.42579 |
| 1960 | 0.025 | 0.07 | 0.163 | 0.207 | 0.234 | 0.24 | 0.268 | 0.242 | 0.36 | 0.43132 |
| 1961 | 0.025 | 0.07 | 0.148 | 0.206 | 0.235 | 0.232 | 0.259 | 0.274 | 0.281 | 0.39639 |
| 1962 | 0.025 | 0.07 | 0.148 | 0.192 | 0.24 | 0.301 | 0.293 | 0.282 | 0.273 | 0.44136 |
| 1963 | 0.025 | 0.07 | 0.148 | 0.193 | 0.243 | 0.275 | 0.311 | 0.363 | 0.329 | 0.46536 |
| 1964 | 0.025 | 0.07 | 0.159 | 0.214 | 0.24 | 0.291 | 0.305 | 0.306 | 0.365 | 0.47387 |
| 1965 | 0.025 | 0.14 | 0.198 | 0.223 | 0.251 | 0.297 | 0.337 | 0.358 | 0.526 | 0.46044 |
| 1966 | 0.025 | 0.07 | 0.16 | 0.149 | 0.389 | 0.31 | 0.406 | 0.377 | 0.385 | 0.50451 |
| 1967 | 0.025 | 0.177 | 0.164 | 0.235 | 0.242 | 0.399 | 0.362 | 0.283 | 0.381 | 0.45912 |
| 1968 | 0.025 | 0.122 | 0.171 | 0.248 | 0.312 | 0.28 | 0.629 | 0.416 | 0.41 | 0.48561 |
| 1969 | 0.025 | 0.137 | 0.174 | 0.252 | 0.324 | 0.364 | 0.579 | 0.415 | 0.469 | 0.52107 |
| 1970 | 0.025 | 0.137 | 0.201 | 0.275 | 0.341 | 0.367 | 0.423 | 0.458 | 0.39 | 0.55442 |
| 1971 | 0.034 | 0.148 | 0.213 | 0.313 | 0.361 | 0.41 | 0.432 | 0.474 | 0.483 | 0.53254 |
| 1972 | 0.038 | 0.155 | 0.218 | 0.313 | 0.419 | 0.443 | 0.443 | 0.443 | 0.508 | 0.60178 |
| 1973 | 0.039 | 0.149 | 0.226 | 0.322 | 0.371 | 0.433 | 0.452 | 0.472 | 0.446 | 0.53554 |
| 1974 | 0.035 | 0.146 | 0.218 | 0.329 | 0.408 | 0.429 | 0.499 | 0.565 | 0.542 | 0.61804 |
| 1975 | 0.035 | 0.148 | 0.206 | 0.311 | 0.403 | 0.446 | 0.508 | 0.582 | 0.58 | 0.6501 |
| 1976 | 0.035 | 0.142 | 0.201 | 0.301 | 0.379 | 0.458 | 0.508 | 0.517 | 0.644 | 0.66481 |
| 1977 | 0.035 | 0.147 | 0.202 | 0.291 | 0.365 | 0.409 | 0.478 | 0.487 | 0.531 | 0.64434 |
| 1978 | 0.035 | 0.139 | 0.211 | 0.29 | 0.365 | 0.429 | 0.427 | 0.385 | 0.542 | 0.64441 |
| 1979 | 0.045 | 0.148 | 0.211 | 0.3 | 0.352 | 0.429 | 0.521 | 0.562 | 0.567 | 0.74343 |
| 1980 | 0.039 | 0.157 | 0.2 | 0.304 | 0.345 | 0.394 | 0.489 | 0.537 | 0.579 | 0.64513 |
| 1981 | 0.05 | 0.137 | 0.2 | 0.305 | 0.364 | 0.402 | 0.454 | 0.522 | 0.561 | 0.62226 |
| 1982 | 0.05 | 0.13 | 0.193 | 0.27 | 0.359 | 0.411 | 0.429 | 0.476 | 0.583 | 0.64223 |
| 1983 | 0.05 | 0.14 | 0.2 | 0.285 | 0.329 | 0.435 | 0.464 | 0.483 | 0.51 | 0.63619 |
| 1984 | 0.05 | 0.133 | 0.203 | 0.268 | 0.348 | 0.386 | 0.488 | 0.591 | 0.567 | 0.66346 |
| 1985 | 0.05 | 0.127 | 0.185 | 0.267 | 0.324 | 0.381 | 0.38 | 0.626 | 0.554 | 0.64227 |
| 1986 | 0.05 | 0.133 | 0.191 | 0.278 | 0.345 | 0.423 | 0.495 | 0.487 | 0.587 | 0.68625 |
| 1987 | 0.05 | 0.154 | 0.191 | 0.262 | 0.357 | 0.381 | 0.406 | 0.454 | 0.332 | 0.61971 |
| 1988 | 0.05 | 0.133 | 0.193 | 0.26 | 0.335 | 0.409 | 0.417 | 0.474 | 0.486 | 0.65433 |
| 1989 | 0.05 | 0.133 | 0.195 | 0.29 | 0.35 | 0.34 | 0.411 | 0.475 | 0.419 | 0.59444 |
| 1990 | 0.05 | 0.148 | 0.203 | 0.294 | 0.357 | 0.447 | 0.399 | 0.494 | 0.481 | 0.65279 |
| 1991 | 0.05 | 0.139 | 0.184 | 0.254 | 0.301 | 0.413 | 0.447 | 0.522 | 0.548 | 0.57344 |
| 1992 | 0.05 | 0.156 | 0.194 | 0.257 | 0.307 | 0.398 | 0.406 | 0.472 | 0.5 | 0.54009 |
| 1993 | 0.05 | 0.128 | 0.184 | 0.229 | 0.265 | 0.293 | 0.344 | 0.482 | 0.437 | 0.58327 |
| 1994 | 0.05 | 0.143 | 0.174 | 0.209 | 0.257 | 0.326 | 0.349 | 0.402 | 0.494 | 0.45895 |
| 1995 | 0.05 | 0.151 | 0.179 | 0.24 | 0.253 | 0.321 | 0.365 | 0.357 | 0.545 | 0.54526 |
| 1996 | 0.05 | 0.147 | 0.178 | 0.208 | 0.274 | 0.268 | 0.321 | 0.375 | 0.402 | 0.54643 |
| 1997 | 0.05 | 0.15 | 0.19 | 0.225 | 0.252 | 0.303 | 0.319 | 0.325 | 0.36 | 0.42402 |
| 1998 | 0.05 | 0.14 | 0.173 | 0.234 | 0.267 | 0.281 | 0.328 | 0.273 | 0.336 | 0.4546 |


|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 0.05 | 0.131 | 0.187 | 0.216 | 0.259 | 0.296 | 0.34 | 0.322 | 0.369 | 0.46388 |
| 2000 | 0.05 | 0.139 | 0.185 | 0.226 | 0.264 | 0.275 | 0.287 | 0.337 | 0.391 | 0.3763 |
| 2001 | 0.05 | 0.144 | 0.185 | 0.223 | 0.263 | 0.319 | 0.327 | 0.421 | 0.41 | 0.53023 |
| 2002 | 0.05 | 0.145 | 0.197 | 0.245 | 0.267 | 0.267 | 0.299 | 0.308 | 0.435 | 0.43536 |
| 2003 | 0.05 | 0.146 | 0.194 | 0.24 | 0.256 | 0.288 | 0.33 | 0.312 | 0.509 | 0.46973 |
| 2004 | 0.05 | 0.137 | 0.195 | 0.24 | 0.245 | 0.305 | 0.316 | 0.448 | 0.356 | 0.60138 |
| 2005 | 0.05 | 0.15 | 0.189 | 0.234 | 0.237 | 0.258 | 0.276 | 0.396 | 0.369 | 0.42863 |
| 2006 | 0.05 | 0.148 | 0.197 | 0.25 | 0.27 | 0.319 | 0.286 | 0.341 | 0.409 | 0.45521 |
| 2007 | 0.05 | 0.152 | 0.179 | 0.216 | 0.242 | 0.245 | 0.275 | 0.252 | 0.257 | 0.36401 |
| 2008 | 0.05 | 0.154 | 0.198 | 0.212 | 0.239 | 0.302 | 0.282 | 0.231 | 0.274 | 0.40044 |
| 2009 | 0.05 | 0.142 | 0.185 | 0.232 | 0.255 | 0.279 | 0.283 | 0.333 | 0.302 | 0.39017 |
| 2010 | 0.05 | 0.149 | 0.2 | 0.23 | 0.272 | 0.307 | 0.336 | 0.336 | 0.361 | 0.41003 |
| 2011 | 0.05 | 0.141 | 0.179 | 0.223 | 0.261 | 0.276 | 0.32 | 0.36 | 0.444 | 0.39082 |
| 2012 | 0.025 | 0.058 | 0.144 | 0.205 | 0.23 | 0.209 | 0.251 | 0.235 | 0.334 | 0.223 |
| 2013 | 0.034 | 0.068 | 0.117 | 0.186 | 0.254 | 0.258 | 0.309 | 0.241 | 0.325 | 0.562 |
| 2014 | 0.022 | 0.079 | 0.136 | 0.188 | 0.212 | 0.227 | 0.228 | 0.29 | 0.343 | 0.603 |
| 2015 | 0.07 | 0.075 | 0.142 | 0.148 | 0.227 | 0.244 | 0.263 | 0.288 | 0.37 | 0.38939 |

Table 10.2.7. North Sea sole. Natural mortality at age and maturity ate age

|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Natural mortality | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| Maturity | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 10.2.8. North Sea sole. Survey tuning indices

| BTS-ISIS | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 7.031 | 7.121 | 3.695 | 1.654 | 0.688 | 0.276 | 0 | 0 | 0 |
| 1986 | 7.168 | 5.183 | 1.596 | 0.987 | 0.623 | 0.171 | 0.158 | 0 | 0.018 |
| 1987 | 6.973 | 12.548 | 1.834 | 0.563 | 0.583 | 0.222 | 0.228 | 0.058 | 0 |
| 1988 | 83.111 | 12.512 | 2.684 | 1.032 | 0.123 | 0.149 | 0.132 | 0.103 | 0.014 |
| 1989 | 9.015 | 68.084 | 4.191 | 4.096 | 0.677 | 0.128 | 0.242 | 0 | 0.051 |
| 1990 | 37.839 | 24.487 | 21.789 | 0.778 | 1.081 | 0.77 | 0.12 | 0.115 | 0.025 |
| 1991 | 4.035 | 28.841 | 6.872 | 6.453 | 0.136 | 0.135 | 0.063 | 0.045 | 0.013 |
| 1992 | 81.625 | 22.284 | 10.449 | 2.529 | 3.018 | 0.09 | 0.162 | 0.078 | 0.02 |
| 1993 | 6.35 | 42.345 | 1.338 | 5.516 | 3.371 | 6.199 | 0.023 | 0.084 | 0.053 |
| 1994 | 7.66 | 7.121 | 19.743 | 0.124 | 1.636 | 0.088 | 0.983 | 0.009 | 0 |
| 1995 | 28.125 | 8.458 | 6.268 | 5.129 | 0.363 | 0.805 | 0.316 | 0.734 | 0.039 |
| 1996 | 3.975 | 7.634 | 1.955 | 1.785 | 2.586 | 0.326 | 0.393 | 0.052 | 0.264 |
| 1997 | 169.343 | 4.919 | 2.985 | 0.739 | 0.71 | 0.38 | 0.096 | 0.035 | 0.042 |
| 1998 | 17.108 | 27.422 | 1.862 | 1.242 | 0.073 | 0.015 | 0.391 | 0 | 0 |
| 1999 | 11.96 | 18.363 | 15.783 | 0.584 | 1.92 | 0.31 | 0.218 | 0.604 | 0.003 |
| 2000 | 14.594 | 6.144 | 4.045 | 1.483 | 0.263 | 0.141 | 0.06 | 0.007 | 0.15 |
| 2001 | 7.998 | 9.963 | 2.156 | 1.564 | 0.684 | 0.074 | 0.037 | 0.028 | 0 |
| 2002 | 20.989 | 4.182 | 3.428 | 0.886 | 0.363 | 0.361 | 0.032 | 0.069 | 0 |
| 2003 | 10.507 | 9.947 | 2.459 | 1.67 | 0.36 | 0.187 | 0.319 | 0 | 0.02 |
| 2004 | 4.192 | 4.354 | 3.553 | 0.644 | 0.626 | 0.118 | 0.07 | 0.073 | 0 |
| 2005 | 5.534 | 3.395 | 2.377 | 1.303 | 0.167 | 0.171 | 0.077 | 0.047 | 0 |
| 2006 | 17.089 | 2.332 | 0.278 | 0.709 | 0.479 | 0.151 | 0.088 | 0 | 0.007 |
| 2007 | 7.498 | 19.504 | 1.464 | 0.565 | 0.315 | 0.537 | 0.031 | 0.009 | 0 |
| 2008 | 15.247 | 9.062 | 12.298 | 1.313 | 0.222 | 0.279 | 0.202 | 0.028 | 0.047 |
| 2009 | 15.95 | 4.999 | 2.858 | 4.791 | 0.252 | 0.124 | 0.272 | 0.079 | 0 |
| 2010 | 54.811 | 10.707 | 2.027 | 0.774 | 1.252 | 0.143 | 0.122 | 0.005 | 0.027 |
| 2011 | 26.166 | 17.387 | 4.006 | 1.094 | 0.778 | 0.828 | 0.013 | 0 | 0.141 |
| 2012 | 5.149 | 18.212 | 8.863 | 1.692 | 0.764 | 0.257 | 0.229 | 0.046 | 0 |
| 2013 | 6.844 | 3.558 | 12.566 | 5.385 | 0.871 | 0.197 | 0.105 | 0.078 | 0.019 |
| 2014 | 18.926 | 15.576 | 3.373 | 6.763 | 3.208 | 0.377 | 0.101 | 0.02 | 0 |
| 2015 | 21.099 | 25.601 | 9.66 | 1.294 | 4.576 | 1.502 | 0.419 | 0.122 | 0.15 |


| SNS | 1 | 2 | 3 | 4 | 5 | 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 5410 | 734 | 238 | 35 | 4 | 0 |
| 1971 | 903 | 1831 | 113 | 3 | 28.9 | 0 |
| 1972 | 1455 | 272 | 149 | NA | 28.3 | 0 |
| 1973 | 5587 | 935 | 84 | 37 | 13 | 0 |
| 1974 | 2348 | 361 | 65 | NA | 0 | 4.4 |
| 1975 | 525 | 865 | 177 | 18 | 0 | 17.1 |
| 1976 | 1399 | 74 | 229 | 27 | 5.7 | 0 |
| 1977 | 3743 | 776 | 104 | 43 | 31.7 | 3.9 |
| 1978 | 1548 | 1355 | 294 | 28 | 99.4 | 13.3 |
| 1979 | 94 | 408 | 301 | 78 | 0 | 16.7 |
| 1980 | 4313 | 89 | 109 | 61 | 3.3 | 0 |
| 1981 | 3737 | 1413 | 50 | 20 | 0 | 0 |
| 1982 | 5857 | 1146 | 228 | 7 | 10 | 0 |
| 1983 | 2621 | 1123 | 121 | 40 | 0 | 19.7 |
| 1984 | 2493 | 1100 | 318 | 74 | 8 | 0 |
| 1985 | 3619 | 716 | 167 | 49 | 4.4 | 0 |
| 1986 | 3705 | 458 | 69 | 31 | 16.7 | 0 |
| 1987 | 1948 | 944 | 65 | 21 | 0 | 0 |
| 1988 | 11227 | 594 | 282 | 82 | 10.2 | 15.5 |
| 1989 | 2831 | 5005 | 208 | 53 | 18.2 | 18.6 |
| 1990 | 2856 | 1120 | 914 | 100 | 49.6 | 12.5 |
| 1991 | 1254 | 2529 | 514 | 624 | 27.2 | 35.8 |
| 1992 | 11114 | 144 | 360 | 195 | 284.8 | 20 |
| 1993 | 1291 | 3420 | 154 | 213 | 0 | 191.7 |
| 1994 | 652 | 498 | 934 | 10 | 59.3 | 0 |
| 1995 | 1362 | 224 | 143 | 411 | 7.1 | 31.1 |
| 1996 | 218 | 349 | 30 | 36 | 90 | 10 |
| 1997 | 10279 | 154 | 190 | 27 | 58.1 | 230 |
| 1998 | 4095 | 3126 | 142 | 99 | 0 | 10 |
| 1999 | 1649 | 972 | 456 | 10 | 20.7 | 0 |
| 2000 | 1639 | 126 | 166 | 118 | 0 | 2 |
| 2001 | 970 | 655 | 107 | 36 | 56.2 | 0 |
| 2002 | 7548 | 379 | 195 | NA | 30.8 | 19.2 |
| 2003 | NA | NA | NA | NA | NA | NA |
| 2004 | 1370 | 624 | 393 | 69 | 53.1 | 7.5 |
| 2005 | 568 | 163 | 124 | NA | 21.3 | 6.7 |
| 2006 | 2726 | 117 | 25 | 30 | 0 | 0 |
| 2007 | 849 | 911 | 33 | 40 | 14.4 | 0 |
| 2008 | 1259 | 259 | 325 | NA | 10 | 0 |
| 2009 | 1932 | 344 | 62 | 103 | 0 | 0 |
| 2010 | 2637 | 237 | 67 | 42 | 23.2 | 0 |
| 2011 | 1248 | 884 | 211 | 112 | 0 | 38 |
| 2012 | NA | NA | NA | NA | NA | NA |
| 2013 | 967 | 427 | 491 | 179 | 50.8 | 7.6 |
| 2014 | 2849 | 448 | 45 | 60 | 34 | 0 |
| 2015 | 3192 | 2334 | 138 | 160 | 162 | 151 |


| DFS0 | nl | be | de | combined |
| :---: | :---: | :---: | :---: | :---: |
| 1970 | 21.56 |  |  |  |
| 1971 | 20.35 |  |  |  |
| 1972 | 0.76 |  |  |  |
| 1973 | 6.52 |  |  |  |
| 1974 | 1.06 |  | 0.21 |  |
| 1975 | 9.65 |  | 3.79 |  |
| 1976 | 4.23 |  | 0.55 |  |
| 1977 | 1.12 |  | 2.80 |  |
| 1978 | 5.80 |  | 3.10 |  |
| 1979 | 12.76 |  | 1.33 |  |
| 1980 | 26.17 |  | 3.56 |  |
| 1981 | 15.61 |  | 2.10 |  |
| 1982 | 12.75 |  | 1.11 |  |
| 1983 | 4.31 | 2.67 | 2.14 |  |
| 1984 | 7.27 | 5.40 | 1.14 |  |
| 1985 | 12.03 | 16.98 | 0.03 |  |
| 1986 | 4.41 | 2.56 | 0.31 |  |
| 1987 | 30.82 | 2.29 | 1.27 |  |
| 1988 | 1.67 | 0.70 | 3.17 |  |
| 1989 | 3.02 | 1.00 | 0.43 |  |
| 1990 | 0.44 | 0.36 | 0.23 | 6.38 |
| 1991 | 14.52 | 2.17 | 0.87 | 167.56 |
| 1992 | 0.76 | 0.16 | 0.19 | 9.27 |
| 1993 | 1.26 | 0.45 | 0.12 | 15.32 |
| 1994 | 1.82 | 0.69 | 0.15 | 22.06 |
| 1995 | 0.28 | 1.57 | 0.09 | 7.06 |
| 1996 | 2.45 | 4.95 | 0.55 | 40.27 |
| 1997 | 2.14 | 1.40 | 0.03 | 26.94 |
| 1998 | 1.26 | 3.48 | 0.18 |  |
| 1999 | 1.34 | 2.31 | 0.10 |  |
| 2000 | 0.72 | 0.53 | 0.12 | 9.50 |
| 2001 | 2.65 | 9.45 | 0.05 | 51.42 |
| 2002 | 2.43 | 13.39 | 0.18 | 58.58 |
| 2003 | 0.62 | 1.50 | 0.10 | 10.61 |
| 2004 | 0.59 | 10.52 | 0.05 | 31.25 |
| 2005 | 2.24 | 5.66 | 0.99 | 40.99 |
| 2006 | 1.04 | 0.34 | 0.12 | 12.57 |
| 2007 | 0.86 | 1.74 | 0.05 | 13.73 |
| 2008 | 0.97 | 0.43 | 0.02 | 11.77 |
| 2009 | 1.22 | 5.52 | 0.31 | 27.33 |
| 2010 | 2.24 | 7.72 | 0.024 | 42.86 |
| 2011 | 0.98 | 0.48 | 0.07 | 12.13 |
| 2012 | 0.92 | 0.43 | 0.05 | 11.23 |
| 2013 | 3.46 | 1.94 | 0.72 | 44.82 |
| 2014 | 1.98 | 0.69 | 0.07 | 23.62 |
| 2015 | $0.56$ | $0.46$ | 0.05 | 7.45 |

Table 10.3.1. North Sea sole. Assessment summary

| Year | Recruitment | SSB | Landings | Discards | Fbar |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 1 |  |  |  | $\begin{aligned} & \text { Ages 2- } \\ & 6 \end{aligned}$ |
|  | thousands | tonnes | tonnes | tonnes |  |
| 1957 | 133173 | 62928 | 13181 | 739 | 0.213 |
| 1958 | 120070 | 65375 | 13235 | 701 | 0.216 |
| 1959 | 446825 | 68284 | 15479 | 937 | 0.228 |
| 1960 | 41799 | 69492 | 17466 | 1260 | 0.257 |
| 1961 | 68261 | 103720 | 27981 | 2777 | 0.305 |
| 1962 | 11057 | 87064 | 28290 | 1481 | 0.347 |
| 1963 | 12754 | 70159 | 22572 | 778 | 0.338 |
| 1964 | 611539 | 51556 | 14772 | 280 | 0.291 |
| 1965 | 151219 | 41217 | 10450 | 895 | 0.265 |
| 1966 | 55092 | 109817 | 32060 | 4723 | 0.297 |
| 1967 | 85651 | 106122 | 32749 | 2668 | 0.389 |
| 1968 | 120554 | 91237 | 34733 | 1819 | 0.501 |
| 1969 | 86543 | 69397 | 23590 | 2116 | 0.542 |
| 1970 | 203014 | 63714 | 21091 | 2200 | 0.522 |
| 1971 | 57000 | 56001 | 21295 | 2700 | 0.504 |
| 1972 | 110647 | 65080 | 25195 | 2221 | 0.525 |
| 1973 | 146093 | 47830 | 20773 | 1631 | 0.556 |
| 1974 | 122131 | 46208 | 20616 | 2120 | 0.551 |
| 1975 | 59092 | 46806 | 19338 | 2178 | 0.505 |
| 1976 | 138917 | 45552 | 16824 | 1631 | 0.464 |
| 1977 | 172911 | 37143 | 15510 | 1658 | 0.458 |
| 1978 | 63381 | 43159 | 18781 | 2233 | 0.483 |
| 1979 | 17880 | 52765 | 23045 | 2133 | 0.517 |
| 1980 | 181294 | 39168 | 16474 | 1040 | 0.531 |
| 1981 | 218640 | 25990 | 13339 | 1740 | 0.536 |
| 1982 | 204054 | 37887 | 20355 | 3565 | 0.558 |
| 1983 | 203846 | 49563 | 25889 | 4071 | 0.604 |
| 1984 | 95271 | 51530 | 27521 | 3874 | 0.646 |
| 1985 | 111974 | 47927 | 25123 | 2846 | 0.644 |
| 1986 | 163489 | 38037 | 18051 | 1841 | 0.593 |
| 1987 | 84106 | 34507 | 15870 | 1950 | 0.534 |
| 1988 | 686583 | 42206 | 17661 | 1996 | 0.499 |
| 1989 | 135009 | 38607 | 21383 | 3764 | 0.489 |
| 1990 | 247976 | 125120 | 48929 | 6398 | 0.491 |
| 1991 | 88638 | 91008 | 38715 | 3537 | 0.494 |
| 1992 | 442592 | 88799 | 34069 | 2956 | 0.506 |
| 1993 | 88109 | 59020 | 29598 | 3671 | 0.534 |
| 1994 | 64572 | 85156 | 34753 | 4031 | 0.573 |
| 1995 | 113851 | 64557 | 30630 | 2296 | 0.618 |
| 1996 | 76031 | 38480 | 20241 | 1749 | 0.654 |


| Year | Recruitment | SSB | Landings | Discards | Fbar |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1997 | 326152 | 32375 | 16388 | 2045 | 0.666 |
| 1998 | 150896 | 24075 | 18008 | 3556 | 0.662 |
| 1999 | 116975 | 48060 | 24942 | 3937 | 0.647 |
| 2000 | 140444 | 40595 | 20461 | 2610 | 0.627 |
| 2001 | 72200 | 34640 | 18453 | 2247 | 0.612 |
| 2002 | 217842 | 35489 | 16617 | 1817 | 0.603 |
| 2003 | 96079 | 27520 | 16691 | 3007 | 0.596 |
| 2004 | 52786 | 39574 | 17420 | 2800 | 0.575 |
| 2005 | 51243 | 31191 | 12936 | 1535 | 0.53 |
| 2006 | 188203 | 25213 | 10577 | 1078 | 0.485 |
| 2007 | 69450 | 18415 | 9460 | 1576 | 0.458 |
| 2008 | 79121 | 36219 | 13352 | 1836 | 0.451 |
| 2009 | 103169 | 31630 | 12399 | 1480 | 0.447 |
| 2010 | 217368 | 31029 | 11971 | 1596 | 0.425 |
| 2011 | 228790 | 30767 | 11997 | 2015 | 0.384 |
| 2012 | 58439 | 38452 | 14339 | 2147 | 0.335 |
| 2013 | 132788 | 48012 | 15059 | 1602 | 0.286 |
| 2014 | 240457 | 45163 | 13302 | 1398 | 0.24 |
| 2015 | 194127 | 49142 | 12630 | 1663 | 0.201 |
| 2016 | 59248 | 64312 |  |  |  |
| Average | 150624 | 53168 | 20587 | 2257 | 0.475 |

Table 10.4.1. North Sea sole. Input table for RCT3 analysis

| yearclass | N_Age_1 | N_Age_2 | DFSO | SNSO | SNS 1 | BTS 1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1974 | 59092 | 52591 | NA | 174.4 | 525.4 | NA |
| 1975 | 138917 | 122596 | NA | 577.5 | 1399.4 | NA |
| 1976 | 172911 | 153675 | NA | 464.6 | 3742.9 | NA |
| 1977 | 63381 | 56908 | NA | 1585 | 1547.7 | NA |
| 1978 | 17880 | 16108 | NA | 10370.5 | 93.8 | NA |
| 1979 | 181294 | 163061 | NA | 3922.7 | 4312.9 | NA |
| 1980 | 218640 | 195134 | NA | 5145.8 | 3737.2 | NA |
| 1981 | 204054 | 179807 | NA | 3240.7 | 5856.5 | NA |
| 1982 | 203846 | 179488 | NA | 2147 | 2621.1 | NA |
| 1983 | 95271 | 84715 | NA | 769.1 | 2493.1 | NA |
| 1984 | 111974 | 100418 | NA | 3334 | 3619.4 | 7.031 |
| 1985 | 163489 | 147256 | NA | 2713.4 | 3705.1 | 7.168 |
| 1986 | 84106 | 75889 | NA | 742 | 1947.9 | 6.973 |
| 1987 | 686583 | 619725 | NA | 13610.1 | 11226.7 | 83.111 |
| 1988 | 135009 | 121800 | NA | 522.7 | 2830.7 | 9.015 |
| 1989 | 247976 | 223426 | NA | 1743.4 | 2856.2 | 37.839 |
| 1990 | 88638 | 79701 | 6.38 | 50.8 | 1253.6 | 4.035 |
| 1991 | 442592 | 396755 | 167.56 | 3639.7 | 11114 | 81.625 |
| 1992 | 88109 | 78597 | 9.27 | 302.9 | 1290.8 | 6.35 |
| 1993 | 64572 | 57149 | 15.32 | 231.3 | 651.8 | 7.66 |
| 1994 | 113851 | 99840 | 22.06 | 4692.7 | 1362.1 | 28.125 |
| 1995 | 76031 | 66385 | 7.06 | 1374.9 | 218.4 | 3.975 |
| 1996 | 326152 | 285876 | 40.27 | 2322.3 | 10279.3 | 169.343 |
| 1997 | 150896 | 133132 | 26.94 | 803 | 4094.6 | 17.108 |
| 1998 | 116975 | 103693 | NA | 327.9 | 1648.9 | 11.96 |
| 1999 | 140444 | 124684 | NA | 2187.9 | 1639.2 | 14.594 |
| 2000 | 72200 | 63920 | 9.5 | 70 | 970.3 | 7.998 |
| 2001 | 217842 | 190844 | 51.42 | 8340 | 7547.5 | 20.989 |
| 2002 | 96079 | 82677 | 58.58 | 1127.7 | NA | 10.507 |
| 2003 | 52786 | 45003 | 10.61 | NA | 1369.5 | 4.192 |
| 2004 | 51243 | 44337 | 31.25 | 162 | 568.1 | 5.534 |
| 2005 | 188203 | 165500 | 40.99 | 305 | 2726.4 | 17.089 |
| 2006 | 69450 | 61423 | 12.57 | 16 | 848.6 | 7.498 |
| 2007 | 79121 | 69716 | 13.73 | 466.9 | 1259.1 | 15.247 |
| 2008 | 103169 | 90232 | 11.77 | 754.7 | 1931.6 | 15.95 |
| 2009 | 217368 | 189870 | 27.33 | 2291 | 2636.9 | 54.811 |
| 2010 | 228790 | 201060 | 42.86 | 333.9 | 1248 | 26.166 |
| 2011 | 58439 | 51568 | 12.13 | 136.3 | 226.6 | 5.149 |
| 2012 | NA | 116775 | 11.23 | 144.7 | 967.4 | 6.844 |
| 2013 | NA | NA | 44.82 | 237.3 | 2849 | 18.926 |
| 2014 | NA | NA | 23.62 | 126 | 3192 | 21.099 |
| 2015 | NA | NA | 7.45 | 109.7 | NA | NA |

Table 10.4.2. North Sea sole. RCT3 results for age 1 in 2016 (yearclass 2015)

Analysis by RCT3 ver4.0

Sole

Data for 4 surveys over 42 years : 1974-2015
Regression type $=C$
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as .00
Minimum of 3 points used for regression

Forecast/Hindcast variance correction used.

```
yearclass:2014
    index slope intercept se rsquare n indices prediction se.pred WAP.weights
    DFSO 1.0005 8.590 0.5719 0.5608 20 3.162 11.75 0.6180}00.157
    SNS0 1.0228 4.783 1.4378 0.1860 37 4.836 9.73 1.5307 0.0257
    SNS1 0.7542 6.049 0.4127 0.7425 37 8.068 12.13 0.4316 0.3233
    BTS1 0.7590 9.754 0.3838 0.7476 28 3.049 12.07 0.4066 0.3642
VPA Mean NA NA NA NA 38 NA 11.71 0.6827 0.1292
yearclass:2015
    index slope intercept se rsquare n indices prediction se.pred WAP.weights
    DFS0 1.0005 8.590 0.5719 0.5608 20 2.008 10.599 0.6409 0.48648
    SNS0 1.0228 4.783 1.4378 0.1860 37 4.698 9.588 1.5354 0.08477
    SNS1 0.7542 6.049 0.4127 0.742537 NA NA NA NA
    BTS1 0.7590 9.754 0.3838 0.747628 NA NA NA NA
VPA Mean NA NA NA NA 38 NA 11.710 0.6827 0.42875
        WAP logWAP int.se
yearclass:2014 15229011.93 0.2454
yearclass:2015 5924810.99 0.4470
```

Table 10.4.3. North Sea sole. RCT3 results for age 2 in 2016 (yearclass 2014)

Analysis by RCT3 ver4.0
Sole
Data for 4 surveys over 42 years : 1974-2015
Regression type $=\mathrm{C}$
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as .00
Minimum of 3 points used for regression

Forecast/Hindcast variance correction used.

```
yearclass:2014
    index slope intercept se rsquare n indices prediction se.pred WAP.weights
    DFS0 1.0295 8.410 0.6056 0.5224 21 3.162 11.666 0.6522 0.15029
    SNS0 1.0249 4.700 1.4534 0.1790 38 4.836 9.656 1.5430
    SNS1 0.7575 5.918 0.4180 0.7333 38 8.068 12.030 0.4368 0.33505
    BTS1 0.7719 9.617 0.4052 0.7226 29 3.049 11.971 0.4286 0.34794
VPA Mean NA NA NA NA 39 NA 11.590 0.6760 0.13987
yearclass:2015
    index slope intercept se rsquare n indices prediction se.pred WAP.weights
    DFS0 1.0295 8.410 0.6056 0.5224 21 2.008 10.478 0.6743 0.45774
    SNS0 1.0249 4.700 1.4534 0.1790 38 4.698 9.514 1.5475 0.08691
    SNS1 0.7575 5.918 0.4180 0.733338 NA NA NA NA
    BTS1 0.7719 9.617 0.4052 0.722629 NA NA NA NA
VPA Mean NA NA NA NA 39 NA 11.590 0.6760 0.45535
```

    WAP logWAP int.se
    yearclass:2014 13723311.830 .2528
yearclass:2015 5421310.900 .4562

Table 10.5.1. North Sea sole. Input and assumptions for 2016 to the short term forecast ( $F$ values presented are assuming $F=$ Fsq)

| ssb | f2-6 | $\begin{gathered} \text { f_dis } 1- \\ 3 \end{gathered}$ | $\begin{gathered} \text { f_hc2- } \\ 6 \end{gathered}$ | recruits | landings | discards | catch | TAC |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 64312 | 0.201 | 0.052 | 0.171 | 59248 | 14823 | 1543 | 16366 | 13262 |  |  |  |
| age | year | f | f.disc | f.land | stock.n | catch.wt | landings.wt | discards.wt | stock.wt | mat | M |
| 1 | 2016 | 0.04 | 0.03 | 0.01 | 59248 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 |
| 2 | 2016 | 0.1 | 0.05 | 0.06 | 163431 | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 |
| 3 | 2016 | 0.22 | 0.06 | 0.16 | 165908 | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 |
| 4 | 2016 | 0.26 | 0.03 | 0.23 | 69351 | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 |
| 5 | 2016 | 0.22 | 0.01 | 0.21 | 20511 | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 |
| 6 | 2016 | 0.2 | 0 | 0.2 | 48812 | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 |
| 7 | 2016 | 0.21 | 0 | 0.2 | 27755 | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 |
| 8 | 2016 | 0.21 | 0 | 0.21 | 8087 | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 |
| 9 | 2016 | 0.21 | 0 | 0.21 | 3861 | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 |
| 10 | 2016 | 0.21 | 0 | 0.21 | 5512 | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 |
| 1 | 2017 | 0.04 | 0.03 | 0.01 | 111851 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 |
| 2 | 2017 | 0.1 | 0.05 | 0.06 | NA | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 |
| 3 | 2017 | 0.22 | 0.06 | 0.16 | NA | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 |
| 4 | 2017 | 0.26 | 0.03 | 0.23 | NA | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 |
| 5 | 2017 | 0.22 | 0.01 | 0.21 | NA | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 |
| 6 | 2017 | 0.2 | 0 | 0.2 | NA | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 |
| 7 | 2017 | 0.21 | 0 | 0.2 | NA | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 |
| 8 | 2017 | 0.21 | 0 | 0.21 | NA | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 |
| 9 | 2017 | 0.21 | 0 | 0.21 | NA | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 |
| 10 | 2017 | 0.21 | 0 | 0.21 | NA | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 |
| 1 | 2018 | 0.04 | 0.03 | 0.01 | 111851 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 |
| 2 | 2018 | 0.1 | 0.05 | 0.06 | NA | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 |
| 3 | 2018 | 0.22 | 0.06 | 0.16 | NA | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 |
| 4 | 2018 | 0.26 | 0.03 | 0.23 | NA | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 |
| 5 | 2018 | 0.22 | 0.01 | 0.21 | NA | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 |
| 6 | 2018 | 0.2 | 0 | 0.2 | NA | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 |
| 7 | 2018 | 0.21 | 0 | 0.2 | NA | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 |
| 8 | 2018 | 0.21 | 0 | 0.21 | NA | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 |
| 9 | 2018 | 0.21 | 0 | 0.21 | NA | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 |
| 10 | 2018 | 0.21 | 0 | 0.21 | NA | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 |

Table 10.5.2. North Sea sole. Results from the short term forecast assuming F = Fsq.

| basis | landings | f2-6 | f_hc2-6 | f_dis 1-3 | discards | catch | ssb2017 | ssb2018 | ssb_change | tac_change |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fmp | 16165 | 0.2 | 0.17 | 0.05 | 1232 | 17397 | 74186 | 70778 | -5 | 31 |
| Ftar | 16165 | 0.2 | 0.17 | 0.05 | 1232 | 17397 | 74186 | 70778 | -5 | 31 |
| Fmsy | 16165 | 0.2 | 0.17 | 0.05 | 1232 | 17397 | 74186 | 70778 | -5 | 31 |
| Fmsy_low | 9326 | 0.11 | 0.09 | 0.03 | 707 | 10032 | 74186 | 78026 | 5 | -24 |
| Fmsy_high | 27393 | 0.37 | 0.31 | 0.09 | 2111 | 29505 | 74186 | 58925 | -21 | 122 |
| Fpa | 31449 | 0.44 | 0.37 | 0.11 | 2435 | 33885 | 74186 | 54660 | -26 | 156 |
| Fpasig | 33633 | 0.48 | 0.41 | 0.12 | 2611 | 36245 | 74186 | 52367 | -29 | 173 |
| Flim | 40580 | 0.62 | 0.53 | 0.16 | 3180 | 43760 | 74186 | 45099 | -39 | 230 |
| SSB>Bpa | 48371 | 0.807 | 0.68 | 0.21 | 3836 | 52207 | 74186 | 37000 | -50 | 294 |
| SSB>Blim | 58778 | 1.132 | 0.96 | 0.29 | 4759 | 63537 | 74186 | 26300 | -65 | 379 |
| SSB>MSYBtrig | 48371 | 0.807 | 0.68 | 0.21 | 3836 | 52207 | 74186 | 37000 | -50 | 294 |
| TACsq | 12326 | 0.148 | 0.13 | 0.04 | 936 | 13262 | 74186 | 74844 | 1 | 0 |
| 15\%_TAC_inc | 14173 | 0.173 | 0.15 | 0.04 | 1078 | 15251 | 74186 | 72887 | -2 | 15 |
| 15\%_TAC_dec | 10478 | 0.125 | 0.11 | 0.03 | 795 | 11273 | 74186 | 76803 | 4 | -15 |
| Fsq* ${ }^{*}$ | 0 | 0 | NA | NA | 0 | 0 | 74186 | 87936 | 19 | -100 |
| Fsq* 0.25 | 4378 | 0.05 | 0.04 | 0.01 | 330 | 4709 | 74186 | 83279 | 12 | -64 |
| Fsq* 0.5 | 8604 | 0.101 | 0.09 | 0.03 | 652 | 9256 | 74186 | 78791 | 6 | -30 |
| Fsq* 0.9 | 14777 | 0.181 | 0.15 | 0.05 | 1125 | 15901 | 74186 | 72248 | -3 | 20 |
| Fsq* ${ }^{*}$ | 16237 | 0.201 | 0.17 | 0.05 | 1238 | 17475 | 74186 | 70701 | -5 | 32 |
| Fsq* ${ }^{*} 1$ | 17737 | 0.222 | 0.19 | 0.06 | 1354 | 19091 | 74186 | 69114 | -7 | 44 |
| Fsq* ${ }^{*} .25$ | 19822 | 0.252 | 0.21 | 0.06 | 1516 | 21338 | 74186 | 66911 | -10 | 61 |
| Fsq* 1.5 | 23149 | 0.302 | 0.26 | 0.08 | 1776 | 24925 | 74186 | 63398 | -15 | 88 |
| Fsq* 1.75 | 26300 | 0.352 | 0.3 | 0.09 | 2025 | 28325 | 74186 | 60076 | -19 | 114 |
| Fsq*2 | 29344 | 0.403 | 0.34 | 0.1 | 2267 | 31610 | 74186 | 56873 | -23 | 138 |

## Table 10.5.3. North Sea sole. Detailed STF table by age, assuming F = Fsq, rescaled

| AGE | YEAR | F | FDISC | FLAND | StOCKN | CATCHWT | LANDINGSWT | DISCARDSWT | stockwt | MAT | M | CATCHN | CATCH | LANDINGSN | LANDINGS | DISCARDSN | DISCARDS | SSB | TSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2016 | 0.04 | 0.03 | 0.01 | 59248 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 | 2191 | 163 | 719 | 102 | 1472 | 61 | 0 | 2488 |
| 2 | 2016 | 0.105 | 0.05 | 0.06 | 163431 | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 | 15521 | 2030 | 8425 | 1486 | 7096 | 544 | 0 | 12094 |
| 3 | 2016 | 0.215 | 0.06 | 0.16 | 165908 | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 | 30572 | 5264 | 22707 | 4557 | 7865 | 708 | 21845 | 21845 |
| 4 | 2016 | 0.262 | 0.03 | 0.23 | 69351 | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 | 15234 | 3256 | 13335 | 3076 | 1900 | 180 | 12067 | 12067 |
| 5 | 2016 | 0.222 | 0.01 | 0.21 | 20511 | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 | 3898 | 980 | 3682 | 956 | 216 | 24 | 4738 | 4738 |
| 6 | 2016 | 0.203 | 0 | 0.2 | 48812 | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 | 8545 | 2282 | 8344 | 2261 | 201 | 21 | 11861 | 11861 |
| 7 | 2016 | 0.206 | 0 | 0.2 | 27755 | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 | 4918 | 1386 | 4870 | 1380 | 48 | 6 | 7401 | 7401 |
| 8 | 2016 | 0.209 | 0 | 0.21 | 8087 | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 | 1452 | 410 | 1446 | 410 | 6 | 1 | 2208 | 2208 |
| 9 | 2016 | 0.209 | 0 | 0.21 | 3861 | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 | 693 | 221 | 692 | 221 | 1 | 0 | 1336 | 1336 |
| 10 | 2016 | 0.209 | 0 | 0.21 | 5512 | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 | 990 | 375 | 989 | 375 | 1 | 0 | 2856 | 2856 |
| 1 | 2017 | 0.04 | 0.03 | 0.01 | 111851 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 | 4136 | 307 | 1357 | 192 | 2779 | 115 | 0 | 4698 |
| 2 | 2017 | 0.105 | 0.05 | 0.06 | 51527 | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 | 4894 | 640 | 2656 | 468 | 2237 | 172 | 0 | 3813 |
| 3 | 2017 | 0.215 | 0.06 | 0.16 | 133133 | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 | 24533 | 4224 | 18221 | 3656 | 6311 | 568 | 17529 | 17529 |
| 4 | 2017 | 0.262 | 0.03 | 0.23 | 121103 | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 | 26603 | 5685 | 23285 | 5371 | 3317 | 314 | 21072 | 21072 |
| 5 | 2017 | 0.222 | 0.01 | 0.21 | 48298 | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 | 9179 | 2307 | 8671 | 2252 | 508 | 56 | 11157 | 11157 |
| 6 | 2017 | 0.203 | 0 | 0.2 | 14860 | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 | 2601 | 695 | 2540 | 688 | 61 | 6 | 3611 | 3611 |
| 7 | 2017 | 0.206 | 0 | 0.2 | 36055 | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 | 6389 | 1800 | 6327 | 1793 | 63 | 8 | 9615 | 9615 |
| 8 | 2017 | 0.209 | 0 | 0.21 | 20445 | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 | 3671 | 1038 | 3656 | 1036 | 15 | 2 | 5581 | 5581 |
| 9 | 2017 | 0.209 | 0 | 0.21 | 5939 | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 | 1066 | 340 | 1065 | 340 | 2 | 0 | 2055 | 2055 |
| 10 | 2017 | 0.209 | 0 | 0.21 | 6884 | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 | 1236 | 468 | 1235 | 468 | 1 | 0 | 3567 | 3567 |
| 1 | 2018 | 0.04 | 0.03 | 0.01 | 111851 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 | 4136 | 307 | 1357 | 192 | 2779 | 115 | 0 | 4698 |
| 2 | 2018 | 0.105 | 0.05 | 0.06 | 97275 | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 | 9238 | 1208 | 5015 | 884 | 4223 | 324 | 0 | 7198 |


| AGE | YEAR | F | FDISC | FLAND | STOCKN | CATCHWT | LANDINGSWT | DISCARDSWT | stockwt | MAT | M | CATCHN | CATCH | LANDINGSN | LANDINGS | DISCARDSN | DISCARDS | SSB | TSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 2018 | 0.215 | 0.06 | 0.16 | 41975 | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 | 7735 | 1332 | 5745 | 1153 | 1990 | 179 | 5527 | 5527 |
| 4 | 2018 | 0.262 | 0.03 | 0.23 | 97179 | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 | 21347 | 4562 | 18685 | 4310 | 2662 | 252 | 16909 | 16909 |
| 5 | 2018 | 0.222 | 0.01 | 0.21 | 84339 | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 | 16028 | 4029 | 15141 | 3932 | 887 | 97 | 19482 | 19482 |
| 6 | 2018 | 0.203 | 0 | 0.2 | 34990 | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 | 6126 | 1636 | 5982 | 1621 | 144 | 15 | 8503 | 8503 |
| 7 | 2018 | 0.206 | 0 | 0.2 | 10976 | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 | 1945 | 548 | 1926 | 546 | 19 | 2 | 2927 | 2927 |
| 8 | 2018 | 0.209 | 0 | 0.21 | 26559 | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 | 4769 | 1348 | 4750 | 1346 | 19 | 2 | 7251 | 7251 |
| 9 | 2018 | 0.209 | 0 | 0.21 | 15015 | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 | 2696 | 860 | 2691 | 859 | 5 | 1 | 5195 | 5195 |
| 10 | 2018 | 0.209 | 0 | 0.21 | 9417 | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 | 1691 | 640 | 1690 | 640 | 1 | 0 | 4879 | 4879 |

Table 10.5.4. North Sea sole. Input and assumptions for 2016 to the short term forecast ( $F$ values presented are for TAC=landings)

| ssb | $\begin{gathered} \text { f2- } \\ 6 \end{gathered}$ | $\mathrm{f}_{-1} \text { dis } 1-$ | $\begin{gathered} \text { f_hc2- } \\ 6 \end{gathered}$ | recruits | landings | discards | catch | TAC |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 64312 | 0.16 | 0.041 | 0.136 | 59248 | 12021 | 1247 | 13269 | 13262 |  |  |  |
| age | year | f | f.disc | f.land | stock.n | catch.wt | landings.wt | discards.wt | stock.wt | mat | M |
| 1 | 2016 | 0.03 | 0.02 | 0.01 | 59248 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 |
| 2 | 2016 | 0.08 | 0.04 | 0.05 | 163431 | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 |
| 3 | 2016 | 0.17 | 0.04 | 0.13 | 165908 | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 |
| 4 | 2016 | 0.21 | 0.03 | 0.18 | 69351 | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 |
| 5 | 2016 | 0.18 | 0.01 | 0.17 | 20511 | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 |
| 6 | 2016 | 0.16 | 0 | 0.16 | 48812 | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 |
| 7 | 2016 | 0.16 | 0 | 0.16 | 27755 | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 |
| 8 | 2016 | 0.17 | 0 | 0.17 | 8087 | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 |
| 9 | 2016 | 0.17 | 0 | 0.17 | 3861 | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 |
| 10 | 2016 | 0.17 | 0 | 0.17 | 5512 | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 |
| 1 | 2017 | 0.04 | 0.03 | 0.01 | 111851 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 |
| 2 | 2017 | 0.1 | 0.05 | 0.06 | NA | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 |
| 3 | 2017 | 0.22 | 0.06 | 0.16 | NA | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 |
| 4 | 2017 | 0.26 | 0.03 | 0.23 | NA | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 |
| 5 | 2017 | 0.22 | 0.01 | 0.21 | NA | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 |
| 6 | 2017 | 0.2 | 0 | 0.2 | NA | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 |
| 7 | 2017 | 0.21 | 0 | 0.2 | NA | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 |
| 8 | 2017 | 0.21 | 0 | 0.21 | NA | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 |
| 9 | 2017 | 0.21 | 0 | 0.21 | NA | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 |
| 10 | 2017 | 0.21 | 0 | 0.21 | NA | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 |
| 1 | 2018 | 0.04 | 0.03 | 0.01 | 111851 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 |
| 2 | 2018 | 0.1 | 0.05 | 0.06 | NA | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 |
| 3 | 2018 | 0.22 | 0.06 | 0.16 | NA | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 |
| 4 | 2018 | 0.26 | 0.03 | 0.23 | NA | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 |
| 5 | 2018 | 0.22 | 0.01 | 0.21 | NA | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 |
| 6 | 2018 | 0.2 | 0 | 0.2 | NA | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 |
| 7 | 2018 | 0.21 | 0 | 0.2 | NA | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 |
| 8 | 2018 | 0.21 | 0 | 0.21 | NA | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 |
| 9 | 2018 | 0.21 | 0 | 0.21 | NA | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 |
| 10 | 2018 | 0.21 | 0 | 0.21 | NA | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 |

Table 10.5.5. North Sea sole. Results from the short term forecast assuming landings = TAC.

| basis | landings | f2-6 | f_hc2-6 | f_dis 1-3 | discards | catch | ssb2017 | ssb2018 | ssb_change | tac_change |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fmp | 16800 | 0.2 | 0.17 | 0.05 | 1264 | 18064 | 77202 | 73429 | -5 | 36 |
| Ftar | 16800 | 0.2 | 0.17 | 0.05 | 1264 | 18064 | 77202 | 73429 | -5 | 36 |
| Fmsy | 16800 | 0.2 | 0.17 | 0.05 | 1264 | 18064 | 77202 | 73429 | -5 | 36 |
| Fmsy_low | 9693 | 0.11 | 0.09 | 0.03 | 725 | 10417 | 77202 | 80960 | 5 | -21 |
| Fmsy_high | 28467 | 0.37 | 0.31 | 0.09 | 2165 | 30631 | 77202 | 61113 | -21 | 131 |
| Fpa | 32680 | 0.44 | 0.37 | 0.11 | 2496 | 35177 | 77202 | 56683 | -27 | 165 |
| Fpasig | 34949 | 0.48 | 0.41 | 0.12 | 2677 | 37625 | 77202 | 54302 | -30 | 184 |
| Flim | 42164 | 0.62 | 0.53 | 0.16 | 3258 | 45422 | 77202 | 46753 | -39 | 242 |
| SSB>Bpa | 51550 | 0.84 | 0.71 | 0.22 | 4040 | 55589 | 77202 | 37000 | -52 | 319 |
| SSB>Blim | 61971 | 1.166 | 0.99 | 0.3 | 4955 | 66927 | 77202 | 26300 | -66 | 405 |
| SSB>MSYBtrig | 51550 | 0.84 | 0.71 | 0.22 | 4040 | 55589 | 77202 | 37000 | -52 | 319 |
| TACsq | 12337 | 0.142 | 0.12 | 0.04 | 925 | 13262 | 77202 | 78155 | 1 | 0 |
| 15\%_TAC_inc | 14187 | 0.166 | 0.14 | 0.04 | 1065 | 15251 | 77202 | 76196 | -1 | 15 |
| 15\%_TAC_dec | 10488 | 0.12 | 0.1 | 0.03 | 785 | 11273 | 77202 | 80116 | 4 | -15 |
| Fsq* 0 | 0 | 0 | NA | NA | 0 | 0 | 77202 | 91259 | 18 | -100 |
| Fsq* 0.25 | 3660 | 0.04 | 0.03 | 0.01 | 273 | 3933 | 77202 | 87366 | 13 | -70 |
| Fsq* 0.5 | 7164 | 0.08 | 0.07 | 0.02 | 535 | 7699 | 77202 | 83644 | 8 | -42 |
| Fsq* 0.9 | 12460 | 0.144 | 0.12 | 0.04 | 934 | 13394 | 77202 | 78025 | 1 | 1 |
| Fsq* ${ }^{*}$ | 13727 | 0.16 | 0.14 | 0.04 | 1030 | 14757 | 77202 | 76682 | -1 | 11 |
| Fsq* 1.1 | 14973 | 0.176 | 0.15 | 0.05 | 1124 | 16097 | 77202 | 75363 | -2 | 21 |
| Fsq* ${ }^{*} .25$ | 16800 | 0.2 | 0.17 | 0.05 | 1264 | 18064 | 77202 | 73429 | -5 | 36 |
| Fsq* 1.5 | 19742 | 0.24 | 0.2 | 0.06 | 1489 | 21231 | 77202 | 70317 | -9 | 60 |

Table 10.5.6. North Sea sole. Detailed STF table by age, assuming landings=TAC, rescaled

| AGE | YEAR | F | FDISC | FLAND | stockn | CATCHWT | LANDINGSWT | DISCARDSWT | stockwt | MAT | M | CATCHN | CATCH | LANDINGSN | LANDINGS | DISCARDSN | DISCARDS | SSB | TSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2016 | 0.031 | 0.02 | 0.01 | 59248 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 | 1748 | 130 | 573 | 81 | 1174 | 49 | 0 | 2488 |
| 2 | 2016 | 0.083 | 0.04 | 0.05 | 163431 | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 | 12462 | 1630 | 6764 | 1193 | 5697 | 437 | 0 | 12094 |
| 3 | 2016 | 0.171 | 0.04 | 0.13 | 165908 | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 | 24806 | 4272 | 18424 | 3697 | 6382 | 574 | 21845 | 21845 |
| 4 | 2016 | 0.208 | 0.03 | 0.18 | 69351 | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 | 12416 | 2653 | 10868 | 2507 | 1548 | 147 | 12067 | 12067 |
| 5 | 2016 | 0.177 | 0.01 | 0.17 | 20511 | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 | 3165 | 796 | 2990 | 776 | 175 | 19 | 4738 | 4738 |
| 6 | 2016 | 0.161 | 0 | 0.16 | 48812 | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 | 6926 | 1849 | 6763 | 1833 | 163 | 17 | 11861 | 11861 |
| 7 | 2016 | 0.163 | 0 | 0.16 | 27755 | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 | 3987 | 1123 | 3948 | 1119 | 39 | 5 | 7401 | 7401 |
| 8 | 2016 | 0.166 | 0 | 0.17 | 8087 | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 | 1177 | 333 | 1173 | 332 | 5 | 1 | 2208 | 2208 |
| 9 | 2016 | 0.166 | 0 | 0.17 | 3861 | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 | 562 | 179 | 561 | 179 | 1 | 0 | 1336 | 1336 |
| 10 | 2016 | 0.166 | 0 | 0.17 | 5512 | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 | 803 | 304 | 802 | 304 | 1 | 0 | 2856 | 2856 |
| 1 | 2017 | 0.04 | 0.03 | 0.01 | 111851 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 | 4136 | 307 | 1357 | 192 | 2779 | 115 | 0 | 4698 |
| 2 | 2017 | 0.105 | 0.05 | 0.06 | 51948 | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 | 4934 | 645 | 2678 | 472 | 2255 | 173 | 0 | 3844 |
| 3 | 2017 | 0.215 | 0.06 | 0.16 | 136038 | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 | 25068 | 4317 | 18619 | 3736 | 6449 | 580 | 17912 | 17912 |
| 4 | 2017 | 0.262 | 0.03 | 0.23 | 126567 | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 | 27803 | 5942 | 24336 | 5614 | 3467 | 328 | 22023 | 22023 |
| 5 | 2017 | 0.222 | 0.01 | 0.21 | 50967 | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 | 9686 | 2435 | 9150 | 2376 | 536 | 59 | 11773 | 11773 |
| 6 | 2017 | 0.203 | 0 | 0.2 | 15554 | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 | 2723 | 727 | 2659 | 721 | 64 | 7 | 3780 | 3780 |
| 7 | 2017 | 0.206 | 0 | 0.2 | 37590 | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 | 6661 | 1877 | 6596 | 1869 | 65 | 8 | 10024 | 10024 |
| 8 | 2017 | 0.209 | 0 | 0.21 | 21327 | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 | 3829 | 1082 | 3814 | 1081 | 16 | 2 | 5822 | 5822 |
| 9 | 2017 | 0.209 | 0 | 0.21 | 6199 | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 | 1113 | 355 | 1111 | 355 | 2 | 0 | 2145 | 2145 |
| 10 | 2017 | 0.209 | 0 | 0.21 | 7185 | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 | 1290 | 488 | 1289 | 488 | 1 | 0 | 3723 | 3723 |
| 1 | 2018 | 0.04 | 0.03 | 0.01 | 111851 | 0.07 | 0.14 | 0.04 | 0.04 | 0 | 0.1 | 4136 | 307 | 1357 | 192 | 2779 | 115 | 0 | 4698 |
| 2 | 2018 | 0.105 | 0.05 | 0.06 | 97275 | 0.13 | 0.18 | 0.08 | 0.07 | 0 | 0.1 | 9238 | 1208 | 5015 | 884 | 4223 | 324 | 0 | 7198 |


| 3 | 2018 | 0.215 | 0.06 | 0.16 | 42318 | 0.17 | 0.2 | 0.09 | 0.13 | 1 | 0.1 | 7798 | 1343 | 5792 | 1162 | 2006 | 181 | 5572 | 5572 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 2018 | 0.262 | 0.03 | 0.23 | 99299 | 0.21 | 0.23 | 0.09 | 0.17 | 1 | 0.1 | 21813 | 4662 | 19093 | 4404 | 2720 | 257 | 17278 | 17278 |
| 5 | 2018 | 0.222 | 0.01 | 0.21 | 88144 | 0.25 | 0.26 | 0.11 | 0.23 | 1 | 0.1 | 16752 | 4210 | 15825 | 4109 | 927 | 101 | 20361 | 20361 |
| 6 | 2018 | 0.203 | 0 | 0.2 | 36924 | 0.27 | 0.27 | 0.1 | 0.24 | 1 | 0.1 | 6464 | 1726 | 6312 | 1711 | 152 | 16 | 8973 | 8973 |
| 7 | 2018 | 0.206 | 0 | 0.2 | 11489 | 0.28 | 0.28 | 0.12 | 0.27 | 1 | 0.1 | 2036 | 574 | 2016 | 571 | 20 | 2 | 3064 | 3064 |
| 8 | 2018 | 0.209 | 0 | 0.21 | 27690 | 0.28 | 0.28 | 0.11 | 0.27 | 1 | 0.1 | 4972 | 1405 | 4952 | 1403 | 20 | 2 | 7559 | 7559 |
| 9 | 2018 | 0.209 | 0 | 0.21 | 15663 | 0.32 | 0.32 | 0.14 | 0.35 | 1 | 0.1 | 2812 | 897 | 2808 | 897 | 5 | 1 | 5419 | 5419 |
| 10 | 2018 | 0.209 | 0 | 0.21 | 9830 | 0.38 | 0.38 | 0.07 | 0.52 | 1 | 0.1 | 1765 | 668 | 1764 | 668 | 1 | 0 | 5093 | 5093 |



Figure 10.2.1. North Sea sole. Time series of landings (reported to ICES) (1957 - present)


Figure 10.2.2. North Sea sole. Time series of discards (reported to ICES) (2002 - present)


Figure 10.2.3. North Sea sole. Data upload in Intercatch: landings \% by country by metier (top); discards in weight (kg) by country by metier (bottom). Sampled and unsampled refers to availability of age-composition information


Figure 10.2. North Sea sole. Landings and discards numbers-at-age

## Landings weights



Discard weights


## Stock weights



Discard weights


## Stock weights from 2002



Figure 10.2.5. North Sea sole. Landing, discard, and mean stock weights at age for the whole time series (top), and for the most recent years (only discard and mean stock weights, 2002 - present)


Figure 10.2.7.1. North Sea sole. Standardized survey tuning indices. BTS-Isis (red), SNS (black)


Figure 10.2.7.1. North Sea sole. Correlation plots for both tuning indices


Figure 10.3.1. North Sea sole. Assessment summary WGNSSK 2016


Figure 10.3.2. North Sea sole. Retrospective performance of assessment summary


Relative contribution of yearclass
Relative contribution of yearclass


Figure 10.5.1. North Sea sole. Pieplots showing relative contribution of intermediate year assumptions for both $\mathrm{F}=\mathrm{Fsq}$ scenario


Figure 10.5.1. North Sea sole. Pieplots showing relative contribution of intermediate year assumptions for both $\mathrm{F}=\mathrm{Ftac}=$ catch scenario


Figure 10.8.1. North Sea sole. Modelled landings (black line) versus observed landings (green area)


Figure 10.8.2. North Sea sole. Modelled catch (black line) versus observed catches (orange area)


Figure 10.8.3. North Sea sole. Modelled discards (black line) versus observed discards (pink area)


Figure 10.8.4. North Sea sole. Landings residuals

Discard residuals


Figure 10.8.5. North Sea sole. Discard residuals


Figure 10.8.6. North Sea sole. Sigmas of different input time series

## 11 Saithe (Pollachius virens) in Subarea 4, 6 and Division 3.a (North Sea, Rockall and West of Scotland, Skagerrak and Kattegat)

The assessment of saithe in Division 3.a and Subareas 4 and 6 is presented as a benchmarked assessment based on the revised assessment protocol specified by the 2016 meeting of WKNSEA (ICES-WKNSEA 2016) and a further revision, put forward after the WGNSSK 2016 meeting to provide a solution to the uncertainty in the assessment and forecast due to the highly uncertain and fluctuating survey indices. The forecast for this stock was re-run in the Autumn because new information from the IBTS Q3 survey triggered the re-opening criterion - the updated forecast is provided in Annex 04.

### 11.1 General

### 11.1.1 Stock definition

A summary of available information on stock definition can be found in the Stock Annex.

### 11.1.2 Ecosystem aspects

No new information on ecosystem aspects was presented at WGNSSK in 2016. A summary of available information, prepared during WKBENCH 2011, can be found in the Stock Annex. No ecosystem aspects were discussed during WKNSEA 2016.

### 11.1.3 Fisheries

A general description of the fishery (along with its historical development) is presented in the Stock Annex.

Saithe are predominantly taken in the trawler fisheries by Norway, Germany, and France. Changes in the fishing pattern of these three fleets began in 2009, but all fleets appear to have largely reverted back to their original fishing patterns (see Stock Annex). For the German and Norwegian fleets, this is mainly along the shelf edge in Subarea 4 and Division 3.a, while French fleets fish along the northern shelf and west of Scotland (Subareas 4 and 6). The Scottish fleets also catch a large amount of saithe in Subareas 4 and 6, which is then discarded due to lack of quota. Discards can also be high in a few Danish and Swedish fisheries in the Skagerrak because these fleets do not have quota allocations.

### 11.1.4 ICES Advice

The information in this section is taken from the Advice summary sheet 2015, section 6.3.35.

## Advice for 2016

ICES advises that when the EU-Norway management strategy is applied, catches in 2016 should be no more than 75049 tonnes. If this stock is not under the EU landing obligation in 2016 and discard rates do not change from the average (2012-2014), this implies landings of no more than 68601 tonnes.

## Management plan

Since SSB is marginally below 200000 tonnes in 2015, paragraph 3 of the EU-Norway management strategy applies, resulting in an F of 0.298 .

### 11.2 Management

In 2012, an EU-Norway request was made to ICES on options to revise the long-term management strategy for saithe (ICES, 2012). Based upon the evaluations, the EU and Norway agreed to keep the existing management strategy. Because the long-term performance was not clear, ICES advised that the strategy should be re-evaluated within four years (i.e. no later than 2016) and revised if necessary.

In 2013, the effects of interannual quota flexibility in the management strategy for saithe were evaluated (ICES, 2013). ICES concluded that the management strategy evaluated is robust to inclusion of interannual quota flexibility in terms of the probability of the stock biomass falling below Blim. This conclusion is conditional on the interannual quota flexibility being suspended when the stock is estimated to be outside safe biological limits. SSB was estimated to be 199270 tonnes for 2015, which was below $B_{\text {pa }}$ (200 000 tonnes).

Changes to the stock assessment and reference points in 2016 imply a need to re-evaluate the management plan in order to ascertain if it can still be considered precautionary under the new stock perception. Until such an evaluation is conducted, advice will follow protocol, i.e., given according to the ICES MSY approach.

### 11.3 Data available

### 11.3.1 Catch

Official landings data for each country participating in the fishery are presented in Table 11.3.1, together with the corresponding WG estimates and the agreed international quota ("total allowable catch" or TAC). During WKNSEA 2016, catch data were updated in InterCatch for the years 2002-2014. Figures 11.3.1 to 11.3.5 and Tables 11.3.2 to 11.3.4 summarise the proportion of landings and discards, for which samples have been provided. Although a large number of fleets do not provide samples for the landings, these do not contribute a large proportion of the catch; $86 \%$ of the landings have been sampled. The amount of samples taken, especially in the targeted trawl fisheries, is an issue (see ICES-WKNSEA 2016). Age compositions for the remaining landings have been determined by averaging within an area (Division 3.a or combined Subareas $4 / 6)$ and a quarter, similar to previous years. This is because the fleets, particularly the target trawl fishery, are targeted the spawning fish in the first two quarters, while a wider range of age classes is captured in the latter part of the year. Discard observations are not available for the fleets landing the vast majority of saithe (Figure 11.3.5). While Norway has a no discarding policy, discarding is not monitored and discard information is not collected. Norwegian discards for the trawler fleet were raised using discard information from the French and German trawler fleets (i.e., the targeted fishery), while discards for other fleets (all counties) were raised similar to landings (stratification by quarter and area). Raised discards accounted for $2 \%$ of the total catch (Table 11.3.2). Discards were raised for all previous years (1967-2014) during WKNSEA 2016 (Figure 11.3.6; ICES-WKNSEA 2016). Details can be found in the Stock Annex and relevant benchmark working documents.

The full time series of catch, landings, and discards is summarized in Table 11.3.5 and illustrated in Figure 11.3.7. Catch has been relatively stable from 1990 through 2008 and
then declined. The WG estimates of saithe discarding (as a proportion of total catch) has declined from early 2000. Discard estimates were lowest for the period when the saithe trawler fleet changed its exploitation pattern (2009-2011). Prior to 2002, discards were estimated using a constant discarding rate (age specific, Figure 11.3.6; see benchmark WD documentation for details). high discards in 2002-2003, 2007 and 2012 were due to reported discarding by Scottish fisheries.

### 11.3.2 Age compositions

International catch and discard data was collated and catch-at-age was generated using InterCatch. Age composition in the landings was based on samples, provided by Denmark, France, Scotland, Germany, and Norway, which account for $86 \%$ of the total landings (also see Table 11.3.4).

Total catch-at-age data are given in Table 11.3.6, while catch-at-age data for each catch component are given in Tables 11.3.7 and 11.3.8. Age 3 fish make up a smaller portion of the landings in recent years (Figure 11.3.8). The last strong year class in the catch appears to be the 2009 year class as seen in the discards in 2012 at age 3 and landings in 2013 at age 4 . From 2016 onwards, saithe fishing in the bottom trawl fleet is covered by the EU Landing Obligation.

### 11.3.3 Weight at age

Weight-at-age from the catch and catch components for ages 3-10+ are presented in Tables 11.3.9-11.3.11 and Figure 11.3.9. Catch weights are also used as stock weights. There was a decreasing trend in mean weight for ages 6 and older, but that has stopped or been reversed (Figure 11.3.9). Weights-at-age for ages 3-5 have been relatively stable, with some variation, over the last decade. Discard weights since 2009 appear to be increasing.

### 11.3.4 Maturity and natural mortality

The following maturity ogive, revised during the 2016 benchmark, is used for all years (see Stock Annex for details):

| AGE | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8 +}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Proportion mature | 0.0 | 0.0 | 0.0 | 0.2 | 0.65 | 0.84 | 0.97 | 1.0 |

A natural mortality rate of 0.2 is used for all ages and years.

### 11.3.5 Catch, effort and research vessel data

Indices used in the assessment are include as Table 11.3.12. Data for the Norwegian, French, and German commercial trawler fleets were combined into one standardized CPUE index, which is then tuned to the exploitable biomass (see Stock Annex for details). One fisheries-independent survey index was included for tuning of the assessment; the survey is the IBTS quarter 3, ages 3-8, 1992-2015 ("IBTS-Q3").

### 11.4 Data analyses

The assessment of North Sea saithe was carried out using a state-space stock assessment model (SAM; stockassessment.org). Alternate models were run after the benchmark and are summarized in the Stock Annex. An exploratory model is included,
which uses the combined standardized cpue index and IBTS Q3 index for tuning, but where stock weights equal catch weights only for those age classes where stock weights were estimated to be larger than catch weights (ages 7+; see ICES-WKNSEA 2016).

### 11.4.1 Exploratory survey-based analyses

Numbers-at-age for saithe ages 3 to 8 (IBTS Q3) on the log-scale, linked by cohort is shown in Figure 11.4.1. A strong year effect is apparent in 2007, 2011, and 2013; this is reflected in the sharp increase in age 4 when compared to earlier cohorts. Within-cohort correlations between ages for the survey index is also presented in Figure 11.4.2. The catch numbers correlate poorly between cohorts for ages 3 and 4, but are stronger for subsequent ages.

Trends by age for the IBTS Q3 index are shown in Figure 11.4.3. Abundance of age 3 and 4 is very low in 2014, but have increased again in 2015.

### 11.4.2 Exploratory catch-at-age-based analyses

The catch curves for total catches are shown in Figure 11.4.4. The curves show that age 3 is only partially recruited to the fishery for the latter cohorts (around the mid 1990s), but fully recruited for many of the earlier cohorts. The catch curves in recent years are less steep than for earlier cohorts, which indicates a change in exploitation occurred. This may be partially explained by declines in catches by the Norwegian purse seine fishery, which occurred in the early 1990s; purse seiners mainly target younger fish. The minimum landing size ( 40 cm in the North Sea) changed around this time, which would also cause a change in exploitation.

The outcome of WKNSEA 2016 was to remove the 3 cpue series for the targeted trawl fisheries, partially due to concerns over using information in the catch-at-age matrix in both the cpue and in the catch-at-age and because more weight was given to 3 indices within the former assessment model (artificially higher weighting to the cpue indices). A standardized combined cpue index was created for the French, German, and Norwegian trawl fleet targeting saithe, which was then tuned to the exploitable biomass, removing the need to use the information in the catch-at-age matrix twice (see WD 2 from WKNSEA 2016 for details). This index is given in Table 11.3.13 and plotted in Figure 11.4.5.

### 11.4.3 Assessments

The benchmark assessment (ICES-WKNSEA) was rejected by WGNSSK. The benchmark model included GAM-derived survey indices for IBTS Q1 and IBTS Q3, plus the combined cpue index. It was subsequently shown that the IBTS Q1 survey did not adequate cover the stock because of movement in and out of the survey area, which was unrelated to abundance (see working documents in Annex 08). It was also questioned whether the IBTS Q3 survey was useful or if the signal was over-ridden by noise and uncertainty in the index. An external review (see technical minutes under Annex 09) helped clarify that the IBTS Q3 index did contain useful information, but that the GAM model used to generate the index was contributing to the noise/uncertainty. This was because the GAM model used a constant spatial effect for all years, whereas year effects were in the data (see Stock Annex and WGNSSK additional working documents in Annex 08). Because of this, the standard Q3 index (derived by using ALKs created annually by roundfish area) was used, but extended to include ages $3-8$ (see working documents in Annex 08).

The external review also questioned the use of generating stock weights from survey indices. Stock weights were generally lower than catch weights before age 7, after which, they were generally higher than catch weights. The benchmark group discussed that this was plausible; fish that are larger for a given age would be selected by the fisheries up to a certain age, after which, selection should drop (i.e., selection is towards an "average" sized fish). The external review advocated the replacement of stock weights with catch weights for all ages as there was some concern over how stock weights were generated (may not be representative of the population). An alternate (exploratory) assessment has been run using stock weights, generated from survey information (see WKNSEA WD 7) for ages 3-6, and catch weights for ages 7+ (the ages where stock weights are greater than catch weights).

Settings used in the assessment are given in Table 11.4.1. SOP correction of the catches has been done on all revised catches (2002-current assessment year).

### 11.4.4 Exploratory assessment with alternative stock weights

Thirty parameters were estimated in the SAM model; the negative log-likelihood value was 360.27. Estimated catchabilities for the Q3 index were higher than the cpue index (Q3 range 0.031 to 0.091 ; cpue 0.004 ). The correlation from the AR1 autocorrelation, which was the correlation random walks for the fishing mortalities, was high (0.798).

Estimated fishing mortality-at-age are given in Table 11.4.2 and illustrated in Figure 11.4.6; estimated population numbers-at-age are in Table 11.4.3.

The $\log$ catchability residuals from the exploratory assessment are shown in Figure 11.4.7 and the retrospective analysis is in Figure 11.4.8.

The historic stock and fishery trends, including $95 \%$ confidence intervals for the exploratory assessment are in Figure 11.4.9. Because fish aged 3-6 make up a high proportion of the total catch, using a lower average weight as stock weights results in lower SSB than in the final assessment model; in 2015, SSB was $5 \%$ lower. The differences in $\mathrm{F}_{4-7}$ and recruitment were negligible.

### 11.4.5 Final assessment

Settings used in the final assessment are as in Table 11.4.1.
Thirty parameters were estimated in the SAM model; the negative log-likelihood value was 358.89 . Estimated catchabilities for the Q3 index were higher than the cpue index, and not that different from the exploratory model (Q3 range 0.031 to 0.091 ; cpue 0.003 ). The correlation from the AR1 autocorrelation, which was the correlation random walks for the fishing mortalities, was high (0.796).

Estimated fishing mortality-at-age are given in Table 11.4.4 and Figure 11.4.6. Estimated population numbers-at-age are in Table 11.4.5.

The residuals are shown in Figure 11.4.10. After accounting for the correlation between ages within years, the IBTS Q3 residuals show less of a pattern; however, the series is still largely positive at the end of the series, when the series is beginning to show an increase in abundance for most ages. The retrospective analysis shows that F tends to be overestimated, while SSB and recruitment tend to be underestimated (Figure 11.4.11).

The SSB estimates over the entire series for the final model are higher than that for the exploratory model (Figure 11.4.9). This is because stock weights are now heavier for
the younger age classes, despite these fish making up only a proportion of the mature biomass.

### 11.5 Historic Stock Trends

The historic stock and fishery trends from the final assessment are presented in Figure 11.5.1 and Table 11.5.1. Because of the benchmark, historic perception of the stock has changed. Recruitment has been highly variable, but shows an overall decline since the mid 1990s. Recruitment is well below the median for the period 2002-2015, used in the forecast. The decline in SSB reversed in 2010 and SSB is now approaching levels seen in the mid-2000s. The final year estimate of SSB is well above $B_{p a}$ and MSY Btrigger. Fishing mortality has generally declined since the mid 1980s. Currently, fishing mortality is below Fmsy.

### 11.6 Recruitment estimates

Currently, no survey provides an estimate of incoming recruitment. The 2002-2015 median value ( 102 million) used in the short-term forecast is below the estimated recruitment for 2015, but below estimates for 2013 and 2014.

### 11.7 Short-term forecasts

A short-term forecast was carried out based on the final assessment.
Weight-at-age in the stock and catch were the mean values for the last 3 years. The exploitation pattern (selectivity pattern) was chosen as the mean exploitation pattern over the last three years. A TAC constraint for the intermediate year was chosen, i.e., the fishing mortality for 2016 was determined such that the landings in 2015 matched the TAC (which was based on landings without the adjustment). Population numbers-at-age for ages 4 and older in 2015 were survivor estimates, while numbers at age 3 were the median estimate of recruitment for the years 2002-2015. The short-term projection was run in SAM.
The input data for the short term forecast are given in Table 11.7.1.
The management options are given in Table 11.7.2. Assuming that the landings in 2016 are scaled to the TAC in 2016 results in an $F_{2015}$ of 0.24 and a SSB in 2016 of 239561 t . Because reference points were re-estimated after the benchmark, the management plan is no longer valid; therefore the MSY approach is used. Total catch in 2017 is 116605 t , where landings (wanted catch) is 110917 t ; this is a $62 \%$ increase in TAC
The contribution of the 2008-2014 year classes to landings in 2017 are shown in Table 11.7.3. The 2012 and 2013 year classes contribute the most to the forecasts; the 2013 year class was large because it was estimated from the resampled median from 2002-2015.

### 11.8 Medium-term and long-term forecasts

No medium-term or long-term forecasts were carried out.

### 11.9 Biological reference points

The biological reference points were re-estimated following the benchmark in 2016 (Table 11.9.1) and further changes during WGNSSK 2016. Data used in the MSY interval analysis were taken from the FLStock object created from the SAM final assessment model presented above. Data represent the latest assessment input and output data,
reflecting the changes in data and model as decided upon at the benchmark and a second external review.

The interval analysis was based on a segmented regression, where the inflection point was set to Bloss. (107 000; Figure 11.9.1). Recruitment is fairly flat with increasing SSB; this is considered a Type 5 stock recruitment characteristic (distinct plateau, wide range of SSB) and for these stocks, $B_{\text {lim }}$ is recommended to be $B_{\text {loss. }} . B_{\text {lim }}=B_{\text {loss }}=107000$. $B_{\text {pa }}$ was estimated to be 150000 , from the equation $B_{p a}=B_{\lim } * 1.4$.
The alternate equation of $B_{p a}=B_{\lim }{ }^{*} \exp \left(1.645^{*} \sigma\right)$, where $\sigma$ is the standard deviation of $\ln (\mathrm{SSB})$ in the final the assessment year, was not used based on the outcome of WGNSSK 2016. The argument against this was that $\mathrm{B}_{\mathrm{pa}}$, could be overly influenced by the standard deviation for stocks where retrospective patterns are strong, while for stocks that use state-space models, estimating uncertainty that accounts for both the process and observation error is often not straightforward; a more precautionary approach (across stocks) was recommended.

The time period for the interval analysis was truncated to 2003-present. Recruitment per SSB showed signs of a cyclic trend over time (Figure 11.9.2). Whether the low productivity observed in recent years is part of cyclic changes or reflects that the stock has entered a new productivity regime is unknown. A change-point analysis, conducted using the changepoint library (Killick et al. 2016) in R, identified three distinct periods: high recruitment in the first years of the series, mid-level recruitment from 1969-2002, and low recruitment 2003-2015. Although the low recruitment period was used in the interval analysis, Bloss was estimated from the entire time series because no low levels of SSB have been observed during the truncated period.

### 11.10 Estimation of $\mathrm{F}_{\mathrm{MSY}}$

All analyses were conducted with Eqsim. The assessment error in the advisory year ( Fcv ) and the autocorrelation ( $\mathrm{F}_{\mathrm{phi}}$ ) were set to values agreed at WKMSYREF4 for stocks where these uncertainties cannot be estimated (ICES-WKMSYREF4 2016). These values, which are normally derived by comparing F values from the latest assessment with forecasted F values in year-1, could not be estimated for North Sea saithe because the assessment model, input data, and age range for $\mathrm{F}_{\mathrm{bar}}$ were changed during the benchmark. Table 11.10 .1 shows the model and data selection settings.
A landings obligation is soon to be mandatory for the North Sea and Skagerrak. Because of this, adjustments had to be made to the landings weights- and numbers-atage for ages $4+$. Discard numbers-at-age were added to the landings and catch weights-at-age were used for landings weights-at-age.

The results of Eqsim simulations run with and without MSYB trigger $^{\text {are shown in Figures }}$ 11.10.1 and 11.10.2, respectively.

The median $\mathrm{F}_{\mathrm{msy}}$ estimated by Eqsim applying a fixed F harvest strategy was 0.36 (Figure 11.10.3, Table 11.9.1). The upper bound of the FMSY range giving at least $95 \%$ of the maximum yield was estimated to be above both $\mathrm{F}_{\mathrm{p} .05}$ and $\mathrm{F}_{\mathrm{pa}}$, and therefore must be constrained $\mathrm{Fp}_{\mathrm{p} .05}$. $\mathrm{F}_{\mathrm{pa}}$ was estimated as $\mathrm{F}_{\mathrm{lim}} / 1.4$ (again opting for the more precautionary approach when estimating $\mathrm{F}_{\mathrm{pa}}$ ). The median of the SSB estimates at $\mathrm{FmSy}^{\text {was }} 233163$ (Figure 3.8.5.4). Median SSB is also shown in Figure 11.10.3.

Btrigger was set to $B_{p a}$ because $F$ was below FMSY for only the last 4 years. When applying the ICES MSY harvest control rule with a Btrigger of 150000 tonnes, median Fmsy increased to 0.37 . The upper bound must be again restricted to Fp .05 . Median SSB values
are lower than under the constant F scenario because of the higher Fmsy values (Figure 11.10.4).

### 11.11 Quality of the assessment and forecast

Although the benchmark attempted to address some of the issues with the previous year's assessments, several issues still remain.
The poor reliability of the recruitment estimate is a major problem for the saithe assessment. There is no survey that adequately covers the recruiting age class.
The commercial CPUE indices may introduce biases into the assessment if changes in fishing patterns occur, as seen in 2009-2011. There are conflicting signals between the survey and fishable biomass index.
The scientific survey used in the assessment does not cover the entire distribution of the stock, however, it is considered generally representative. The number of observations (trawl stations) with saithe is low and the resulting survey index is uncertain.
The fraction of fish age 3 migrating into the survey area (and the fishery) is low and varying between years with no obvious trend. Observations of saithe at age 3 are not suitable for predicting year class strength. This means that assumed recruitment values are highly uncertain and a substantial portion (30\%) of the advised wanted catch in 2017 is based on the recruitment assumptions for 2016 and 2017.

Status of the Stock
The general perception of the status of the saithe stock is more positive than last year.

### 11.12 Management Considerations

The assessment is sensitive to relatively small changes in the input data. Because this stock suffers from 'poor data', the assessment is relatively uncertain. Recruitment is currently at a low level and it appears the strong recruitment pulses are more sporadic than in the past.

The reported landings have been relatively stable since the early 1990s. Landings have been lower than the TAC since 2002, even with reductions in the TAC were in place 2013-2015. The TAC was taken fully in 2015. Information from fishers' survey (Napier, 2014) has been moved to the Stock Annex.

Bycatch of other demersal fish species does occur in the target trawl fishery for saithe. Saithe is also taken as unintentional bycatch in other fisheries, and discards do occur. Bycatch of saithe in all fisheries in 2015 was estimated to be approximately \% of the official landings; this included estimates of the Norwegian discards.

### 11.12.1 Evaluation of the Management plan

Because reference points were re-estimated after the benchmark, the management plan is no longer valid; therefore the MSY approach is used. The catch option for 2017 based on the EU-Norway management strategy has a lower F than the corresponding FMSY option and is considered precautionary.
The assessment, if run in terms with the management plan, is consistent with the precautionary approach in the short term, conditional on the absence of major changes in the productivity and the absence of measurement and implementation error (ICES Advice 2008, Book 6, Paragraph 6.3.3.3). The EU-Norway management plan was reconsidered in February 2013, but no modification was implemented. It was previously
evaluated by ICES (ICES, 2012a) and considered to be consistent with the precautionary approach in the short term (<4 years). Because the long-term performance was not clear, ICES advised that the strategy should be re-evaluated within four years (i.e. no later than 2016) and revised if necessary.

### 11.13 Criteria for identification of candidate stocks for less frequent assessments

The following criteria were applied to determine whether saithe is a stock that may be subject to a change in the frequency by which it is assessed.

> CRITERIA TO BE USED TO IDENTIFY CANDIDATE STOCKS FOR LESS $$
\text { FREQUENT ASSESSMENT. }
$$

Stocks are considered candidates for biennial assessment if:
The advice for the stock has been 0-catch or equivalent for the latest three advice years.

Stocks are considered candidates for biennial assessment if the following criteria are fulfilled simultaneously:

Life span (i.e. maximum normal age) of the species is larger than 5 years

The stock status in relation to the reference points is according to the MSY criteria F (latest assessment year) <= $1.1 \times$ FMSY OR if FMSY range has been defined: F (latest assessment year) is <= Fupper (upper bound in F range) AND SSB(start of intermediate year) $>=$ MSY Btrigger
The average contribution to the catch in numbers of the recruiting year class in latest 5 years is less than $25 \%$ of the total catch in numbers. Should be calculated as the average over the latest five years of the catch in numbers of first age divided by the total catch in number by year.

The retrospective pattern, based on a seven years peel of Mohn's Rho index, shows that F is consistently underestimated by less than $20 \%$.

```
False
True
True: F=0.301 <= 1.1*FMSY OR
F=0.301 <= MSY Fupper (0.403)
AND SSB2016 (239561) is >= MSY
Btrigger (150 000)
True. The average over the latest five years (in numbers) \(=14 \%\).
Avg. contribution: 2011: 19\%, 2012: 16\%, 2013: \(10 \%\), 2014: \(8 \%\), 2015: 15\%
```

True; $\mathrm{p}<0.01$
Saithe is a stock that meets the criteria for biennial assessment. However, the input data for the assessment is highly uncertain, as shown during the benchmark (ICESWKNSEA 2016) and work following the benchmark. Minor modifications to the input data can greatly change perception of the stock. This, coupled with the uncertainties in the forecast, mean that saithe should continue to be assessed annually.

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Table 11.3.1. Saithe in Subareas 4 and 6 and Division 3.a. Nominal landings (tonnes) of saithe, 20042015, as officially reported to ICES and estimated by the Working Group.

Subarea 4 and Division 3.A

| Country | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 22 | 28 | 15 | 18 | 7 | 27 | 15 | 2 | 1 | 3 | 4 | 6 |
| Denmark | 7991 | 7498 | 7470 | 5443 | 8066 | 8802 | 8018 | 6331 | 5171 | 5691 | 5056 | 4508 |
| Faroe Isl. | 558 | 463 | 60 | 15 | 108 | 841 | 146 | 2 | 8 | 3 | 0 | 0 |
| France | 13628 | 11830 | 16953 | 15083 | 15881 | 7203 | 4582* | 13856* | 14093* | 8475 | 7906 | 11612 |
| Germany | 9589 | 12401 | 14397 | 12791 | 14140 | 13410 | 11193 | 10234 | 8052 | 9687 | 8562 | 7954 |
| Greenland | 403 | 1042 | 924 | 564 | 888 | 927 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ireland | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lithuania | 0 | 149 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Netherlands | 3 | 40 | 28 | 5 | 3 | 16 | 3 | 24 | 34 | 168 | 0 | 64 |
| Norway | 62783 | 68122 | 61318 | 45396 | 61464 | 57708 | 52712 | 46809 | 33288 | 35701 | 37463 | 35691 |
| Poland | 0 | 1100 | 1084 | 1384 | 1407 | 988 | 654 | 584 | 0 | 0 | 0 | 0 |
| Russia | 0 | 35 | 2 | 5 | 5 | 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sweden | 2249 | 2132 | 1745 | 1381 | 1639 | 1363 | 1545 | 1335 | 1306 | 1401 | 1272 | 1157 |
| UK (E/W/NI) | 457 | 960 |  |  |  |  |  |  |  |  | 687 | 8888** |
| UK (Scotland) | 5924 | 6170 | 9128** | 9625** | 11804** | 12584** | 11887** | 10250** | 7287** | 10379** | 7686 |  |
| Total reported | 103608 | 111970 | 113124 | 91710 | 115412 | 103883 | 90755 | 89427 | 69240 | 71508 | 68318 | 69879 |
| Unallocated | -3646 | -427 | 3988 | 1908 | -3979 | 1646 | 4345 | 277 | 645 | 317 | 319 | 726 |
| ICES estimate | 99962 | 111543 | 117112 | 93618 | 111433 | 105529 | 95100 | 89704 | 70510 | 71825 | 68662 | 69153 |
| TAC | 190000 | 145000 | 123250 | 135900 | 135900 | 125934 | 107000 | 93600 | 79320 | 91220 | 77536 | 66006 |

*Preliminary.
**Scotland+E/W/NI combined.

SUBAREA 6

| CounTRY | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ | $\mathbf{2 0 1 0}$ | $\mathbf{2 0 1 1}$ | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ | $\mathbf{2 0 1 4}$ | $\mathbf{2 0 1 5 *}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Denmark | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 20 | 0 |
| Faroe Islands | 34 | 25 | 76 | 32 | 23 | 60 | 24 | 5 | 6 | 25 | 0 | 3 |
| France | 3053 | 3954 | 6092 | 4327 | 4170 | 2102 | 2008 | 2357 | 2612 | 3814 | 2904 | 3484 |
| Germany | 4 | 373 | 532 | 580 | 148 | 298 | 257 | 0 | 9 | 0 | 0 | 0 |
| Ireland | 95 | 168 | 267 | 322 | 288 | 407 | 520 | 359 | 364 | 313 | 128 | 105 |
| Netherlands | 0 | 0 | 3 | 36 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 6 |
| Norway | 16 | 20 | 28 | 377 | 78 | 68 | 121 | 240 | 5 | 715 | 442 | 677 |
| Russia | 6 | 25 | 7 | 2 | 50 | 4 | 2 | 0 | 0 | 0 | 0 | 1 |
| Spain | 2 | 3 | 6 | 3 | 4 | 8 | 18 | 31 | 13 | 21 | 0 | 15 |
| UK (E/W/NI) | 37 | 133 |  |  |  |  |  |  |  |  | 97 |  |
| UK (Scotland) | 1563 | 2922 | $2748^{* *}$ | $1424^{* *}$ | $2955^{* *}$ | $3491^{* *}$ | $3168^{* *}$ | $4500^{* *}$ | $4549^{* *}$ | $3646^{* *}$ | 3191 | $3286^{* *}$ |
| Total reported | 4810 | 7623 | 9759 | 7103 | 7717 | 6438 | 6118 | 7492 | 7558 | 8534 | 6842 | 7577 |
| Unallocated | -296 | -1884 | -1191 | -317 | -483 | 525 | 722 | -92 | -351 | -472 | -60 | -1578 |
| ICES estimate | 4514 | 5739 | 8568 | 6786 | 7234 | 6963 | 6840 | 7400 | 7162 | 8062 | 6831 | 9155 |
| TAC | 20000 | 15044 | 12787 | 14100 | 14100 | 13066 | 11000 | 9570 | 8230 | 9464 | 8045 | 6848 |

*Preliminary.
**Scotland+E/W/NI combined.

Subarea 4, Division 3.a, and Subarea 6

|  | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES estimate | 108418 | 119593 | 125680 | 100404 | 118667 | 112492 | 101940 | 97104 | 77672 | 79887 | 75419 | 78307 |
| TAC | 210000 | 160044 | 136037 | 150000 | 150000 | 139000 | 118000 | 103170 | 87550 | 100684 | 85581 | 72854 |

Table 11.3.2. Saithe in Subareas 4 and 6 and Division 3.a. Proportion of sampling strata for discards imported into InterCatch and proportion of discards raised from averaged discard rates.

| CATCH CATEGORY | Raised Or Imported |  | Weight (TONNES) |
| :--- | :--- | :--- | :--- |
| Piscards | Raised_Discards | 94.9 | 2 |
| Discards | Imported_Data | 4913.8 | 98 |
| Landings | Imported_Data | 76936.3 | 100 |

Table 11.3.3. Saithe in Subareas 4 and 6 and Division 3.a. Proportion of distributions for landings and discards either imported or raised in InterCatch and either sampled or estimated.

| Catch Category | Raised Or Imported | Sampled Or Estimated | Weight (tonnes) | Proportion |
| :---: | :---: | :---: | :---: | :---: |
| Landings | Imported_Data | Sampled_Distribution | 66398.8 | 86 |
| Landings | Imported_Data | Estimated_Distribution | 10537.5 | 14 |
| Discards | Imported_Data | Sampled_Distribution | 4644.6 | 93 |
| Discards | Imported_Data | Estimated_Distribution | 269.2 | 5 |
| Discards | Raised_Discards | Estimated_Distribution | 94.9 | 2 |

Table 11.3.4. Saithe in Subareas 4 and 6 and Division 3.a. Proportion by area of distributions for landings and discards either imported or raised in InterCatch and either sampled or estimated.

| CATCH <br> CATEGORY | RAISED OR <br> IMPORTED |  | SAMPLED OR EstIMATED |  |  |  | AREA | (TONNES) | PROPORTION |
| :--- | :--- | :--- | :--- | :--- | :--- | :---: | :---: | :---: | :---: |
| Landings | Imported_Data | Sampled_Distribution | 6 | 6160 | 82 |  |  |  |  |
| Landings | Imported_Data | Estimated_Distribution | 6 | 1373 | 18 |  |  |  |  |
| Discards | Imported_Data | Sampled_Distribution | 6 | 269 | 66 |  |  |  |  |
| Discards | Imported_Data | Estimated_Distribution | 6 | 134 | 33 |  |  |  |  |
| Discards | Raised_Discards | Estimated_Distribution | 6 | 2 | 0 |  |  |  |  |
| Landings | Imported_Data | Sampled_Distribution | 4 | 59406 | 88 |  |  |  |  |
| Landings | Imported_Data | Estimated_Distribution | 4 | 8404 | 12 |  |  |  |  |
| Discards | Imported_Data | Sampled_Distribution | 4 | 4266 | 98 |  |  |  |  |
| Discards | Raised_Discards | Estimated_Distribution | 4 | 56 | 1 |  |  |  |  |
| Discards | Imported_Data | Estimated_Distribution | 4 | 35 | 1 |  |  |  |  |
| Landings | Imported_Data | Sampled_Distribution | $3 . a N$ | 832 | 52 |  |  |  |  |
| Landings | Imported_Data | Estimated_Distribution | $3 . a N$ | 760 | 48 |  |  |  |  |
| Discards | Imported_Data | Sampled_Distribution | $3 . a N$ | 109 | 44 |  |  |  |  |
| Discards | Imported_Data | Estimated_Distribution | $3 . a N$ | 100 | 41 |  |  |  |  |
| Discards | Raised_Discards | Estimated_Distribution | $3 . a N$ | 37 | 15 |  |  |  |  |

Table 11.3.5. Saithe in Subareas 4 and 6 and Division 3.a. Working Group estimates of catch components by weight ( 000 tonnes).

| Year | Catches | LANDINGS | DISCARDS | Proportion discards |
| :---: | :---: | :---: | :---: | :---: |
| 1967 | 126743 | 113751 | 12992 | 10 |
| 1968 | 109144 | 88326 | 20818 | 19 |
| 1969 | 150301 | 130588 | 19713 | 13 |
| 1970 | 270779 | 234962 | 35817 | 13 |
| 1971 | 309202 | 265381 | 43821 | 14 |
| 1972 | 296444 | 261877 | 34567 | 12 |
| 1973 | 275150 | 242499 | 32651 | 12 |
| 1974 | 337025 | 298351 | 38674 | 11 |
| 1975 | 304619 | 271584 | 33035 | 11 |
| 1976 | 423416 | 343967 | 79449 | 19 |
| 1977 | 239915 | 216395 | 23520 | 10 |
| 1978 | 176868 | 155141 | 21727 | 12 |
| 1979 | 142655 | 128360 | 14295 | 10 |
| 1980 | 145300 | 131908 | 13392 | 9 |
| 1981 | 148249 | 132278 | 15971 | 11 |
| 1982 | 202126 | 174351 | 27775 | 14 |
| 1983 | 203022 | 180044 | 22978 | 11 |
| 1984 | 240557 | 200834 | 39723 | 17 |
| 1985 | 273671 | 220869 | 52802 | 19 |
| 1986 | 232786 | 198596 | 34190 | 15 |
| 1987 | 192391 | 167514 | 24877 | 13 |
| 1988 | 154248 | 135172 | 19076 | 12 |
| 1989 | 124584 | 108877 | 15707 | 13 |
| 1990 | 124419 | 103800 | 20619 | 17 |
| 1991 | 130950 | 108048 | 22902 | 17 |
| 1992 | 115534 | 99742 | 15792 | 14 |
| 1993 | 132610 | 111491 | 21119 | 16 |
| 1994 | 126760 | 109622 | 17138 | 14 |
| 1995 | 141205 | 121810 | 19395 | 14 |
| 1996 | 128925 | 114997 | 13928 | 11 |
| 1997 | 120082 | 107327 | 12755 | 11 |
| 1998 | 117219 | 106123 | 11096 | 9 |
| 1999 | 119652 | 110716 | 8936 | 7 |
| 2000 | 99336 | 91322 | 8014 | 8 |
| 2001 | 106160 | 95042 | 11118 | 10 |
| 2002 | 143580 | 122036 | 21544 | 15 |
| 2003 | 123821 | 112383 | 11438 | 9 |
| 2004 | 115472 | 107384 | 8088 | 7 |
| 2005 | 127069 | 118873 | 8196 | 6 |
| 2006 | 130235 | 121650 | 8585 | 7 |
| 2007 | 111883 | 99470 | 12413 | 11 |
| 2008 | 130207 | 121848 | 8359 | 6 |
| 2009 | 118052 | 113756 | 4296 | 4 |
| 2010 | 107488 | 103004 | 4484 | 4 |
| 2011 | 101960 | 97598 | 4362 | 4 |
| 2012 | 87143 | 77865 | 9278 | 11 |
| 2013 | 88224 | 80447 | 7777 | 9 |
| 2014 | 81830 | 75493 | 6337 | 8 |
| 2015 | $83310$ | 78307 | 5003 | 6 |

Table 11.3.6. Saithe in Subareas 4 and 6 and Division 3.a. Catch numbers (thousands) at age for the age range used in the assessment.

| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 26948 | 19395 | 16672 | 2358 | 1610 | 299 | 203 | 185 |
| 1968 | 36111 | 25387 | 14153 | 6166 | 433 | 247 | 127 | 147 |
| 1969 | 47014 | 21142 | 11869 | 7790 | 5795 | 810 | 642 | 151 |
| 1970 | 57920 | 91668 | 16102 | 12416 | 3932 | 1834 | 326 | 270 |
| 1971 | 108549 | 69105 | 35143 | 4848 | 4290 | 2910 | 1922 | 782 |
| 1972 | 74755 | 79033 | 27178 | 21711 | 3709 | 3014 | 1682 | 1625 |
| 1973 | 84484 | 45078 | 28822 | 16443 | 8511 | 2047 | 1391 | 2407 |
| 1974 | 104086 | 40345 | 15160 | 21179 | 14810 | 5321 | 1514 | 1977 |
| 1975 | 88613 | 30927 | 11077 | 7746 | 13792 | 9577 | 3591 | 2717 |
| 1976 | 323156 | 63447 | 12556 | 6401 | 4016 | 5488 | 3678 | 3528 |
| 1977 | 42701 | 65727 | 15839 | 5620 | 3814 | 3528 | 3909 | 4753 |
| 1978 | 54515 | 32608 | 19389 | 3390 | 1149 | 1057 | 788 | 3522 |
| 1979 | 25395 | 16999 | 12004 | 8906 | 2833 | 750 | 554 | 2112 |
| 1980 | 27203 | 14757 | 9677 | 6878 | 5714 | 1177 | 522 | 2327 |
| 1981 | 40705 | 9971 | 7235 | 3763 | 3368 | 3475 | 674 | 2564 |
| 1982 | 49595 | 48533 | 9848 | 6120 | 2166 | 1489 | 1007 | 1268 |
| 1983 | 43916 | 24637 | 27924 | 5813 | 4942 | 1529 | 1062 | 1342 |
| 1984 | 125848 | 38470 | 13910 | 13320 | 1673 | 1281 | 344 | 653 |
| 1985 | 208401 | 66489 | 14257 | 4878 | 3034 | 698 | 409 | 750 |
| 1986 | 86198 | 109080 | 16302 | 5509 | 2629 | 1490 | 457 | 910 |
| 1987 | 48545 | 116551 | 15019 | 3233 | 1829 | 1269 | 933 | 707 |
| 1988 | 50657 | 31577 | 37919 | 3918 | 1927 | 1130 | 796 | 687 |
| 1989 | 34408 | 36772 | 14156 | 11211 | 1572 | 757 | 430 | 493 |
| 1990 | 63454 | 23416 | 12154 | 4826 | 2803 | 762 | 288 | 368 |
| 1991 | 71710 | 35719 | 8016 | 3669 | 1733 | 976 | 376 | 463 |
| 1992 | 28617 | 40193 | 13691 | 3269 | 1539 | 712 | 531 | 426 |
| 1993 | 58813 | 24905 | 12715 | 3199 | 1583 | 1547 | 835 | 1037 |
| 1994 | 31034 | 48062 | 13992 | 4399 | 957 | 354 | 438 | 803 |
| 1995 | 41461 | 31130 | 15884 | 3864 | 3529 | 690 | 566 | 809 |
| 1996 | 17208 | 46468 | 12653 | 7915 | 3194 | 827 | 215 | 496 |
| 1997 | 23380 | 23077 | 32395 | 3763 | 2666 | 1036 | 299 | 292 |
| 1998 | 16113 | 37088 | 17570 | 16459 | 2253 | 1234 | 581 | 280 |
| 1999 | 14661 | 16588 | 28645 | 8588 | 10169 | 2401 | 914 | 665 |
| 2000 | 10985 | 20680 | 9597 | 12632 | 3190 | 3302 | 657 | 446 |
| 2001 | 24961 | 21100 | 24068 | 3429 | 3621 | 1814 | 1655 | 248 |
| 2002 | 17570 | 37489 | 14736 | 13731 | 2309 | 2544 | 1321 | 1575 |
| 2003 | 28296 | 31752 | 20631 | 6836 | 6855 | 1535 | 2000 | 2042 |
| 2004 | 13642 | 24479 | 15649 | 15220 | 2037 | 2164 | 1300 | 1066 |
| 2005 | 12690 | 15473 | 19060 | 20042 | 7956 | 1628 | 1188 | 1151 |
| 2006 | 17313 | 31972 | 10381 | 11286 | 8395 | 3824 | 1008 | 1281 |
| 2007 | 24614 | 13314 | 20919 | 7175 | 5564 | 3610 | 1218 | 930 |
| 2008 | 7620 | 30911 | 12540 | 14941 | 5088 | 3285 | 3551 | 3118 |
| 2009 | 7438 | 15507 | 14222 | 5847 | 8512 | 2994 | 1519 | 2945 |
| 2010 | 8766 | 9249 | 9440 | 6511 | 2671 | 4773 | 1679 | 2707 |
| 2011 | 12786 | 24269 | 8980 | 3674 | 2867 | 1208 | 1564 | 3877 |
| 2012 | 14334 | 13053 | 16948 | 4075 | 1977 | 1268 | 541 | 2611 |
| 2013 | 7267 | 30318 | 5312 | 7869 | 1890 | 1241 | 616 | 1658 |
| 2014 | 4055 | 14322 | 15195 | 3957 | 4124 | 1040 | 429 | 1389 |
| 2015 | 8369 | 8323 | 14259 | 8254 | 1862 | 1623 | 715 | 977 |

Table 11.3.7. Saithe in Subareas 4 and 6 and Division 3.a. Landings numbers (thousands) at age for the age range used in the assessment.

| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 17330 | 16220 | 15531 | 2303 | 1594 | 292 | 198 | 183 |
| 1968 | 23223 | 21231 | 13184 | 6023 | 429 | 242 | 123 | 145 |
| 1969 | 30235 | 17681 | 11057 | 7609 | 5738 | 791 | 626 | 150 |
| 1970 | 37249 | 76661 | 15000 | 12128 | 3894 | 1792 | 318 | 267 |
| 1971 | 69808 | 57792 | 32737 | 4736 | 4248 | 2843 | 1874 | 774 |
| 1972 | 48075 | 66095 | 25317 | 21207 | 3672 | 2944 | 1641 | 1607 |
| 1973 | 54332 | 37698 | 26849 | 16061 | 8428 | 2000 | 1357 | 2381 |
| 1974 | 66938 | 33740 | 14123 | 20688 | 14666 | 5199 | 1477 | 1955 |
| 1975 | 56987 | 25864 | 10319 | 7566 | 13657 | 9357 | 3501 | 2687 |
| 1976 | 207823 | 53060 | 11696 | 6253 | 3976 | 5362 | 3586 | 3490 |
| 1977 | 27461 | 54967 | 14755 | 5490 | 3777 | 3447 | 3812 | 4701 |
| 1978 | 35059 | 27269 | 18062 | 3312 | 1138 | 1033 | 768 | 3484 |
| 1979 | 16332 | 14216 | 11182 | 8699 | 2805 | 733 | 540 | 2089 |
| 1980 | 17494 | 12341 | 9015 | 6718 | 5658 | 1150 | 509 | 2302 |
| 1981 | 26178 | 8339 | 6739 | 3675 | 3335 | 3396 | 657 | 2536 |
| 1982 | 31895 | 40587 | 9174 | 5978 | 2145 | 1454 | 982 | 1254 |
| 1983 | 28242 | 20604 | 26013 | 5678 | 4893 | 1494 | 1036 | 1327 |
| 1984 | 80933 | 32172 | 12957 | 13011 | 1657 | 1252 | 335 | 646 |
| 1985 | 134024 | 55605 | 13281 | 4765 | 3005 | 682 | 399 | 742 |
| 1986 | 55435 | 91223 | 15186 | 5381 | 2603 | 1456 | 445 | 900 |
| 1987 | 31220 | 97470 | 13990 | 3158 | 1811 | 1240 | 910 | 700 |
| 1988 | 32578 | 26408 | 35323 | 3828 | 1908 | 1104 | 776 | 680 |
| 1989 | 22128 | 30752 | 13187 | 10951 | 1557 | 739 | 419 | 488 |
| 1990 | 40808 | 19583 | 11322 | 4714 | 2776 | 745 | 281 | 364 |
| 1991 | 46117 | 29871 | 7467 | 3583 | 1716 | 953 | 367 | 458 |
| 1992 | 18404 | 33614 | 12753 | 3193 | 1524 | 696 | 518 | 422 |
| 1993 | 37823 | 20828 | 11845 | 3125 | 1568 | 1511 | 814 | 1026 |
| 1994 | 19958 | 40194 | 13034 | 4297 | 947 | 346 | 427 | 794 |
| 1995 | 26664 | 26034 | 14797 | 3774 | 3494 | 674 | 552 | 800 |
| 1996 | 11066 | 38861 | 11786 | 7731 | 3163 | 808 | 210 | 491 |
| 1997 | 15036 | 19299 | 30177 | 3676 | 2640 | 1012 | 291 | 288 |
| 1998 | 10363 | 31017 | 16367 | 16077 | 2231 | 1206 | 567 | 277 |
| 1999 | 9429 | 13872 | 26684 | 8389 | 10070 | 2346 | 891 | 657 |
| 2000 | 7064 | 17295 | 8940 | 12339 | 3159 | 3226 | 641 | 441 |
| 2001 | 16052 | 17646 | 22421 | 3349 | 3586 | 1772 | 1614 | 245 |
| 2002 | 9131 | 31779 | 12286 | 13307 | 2245 | 2220 | 1199 | 1479 |
| 2003 | 13009 | 24646 | 20397 | 6836 | 6855 | 1535 | 2000 | 2042 |
| 2004 | 8037 | 20071 | 15649 | 15220 | 2037 | 2164 | 1300 | 1066 |
| 2005 | 9191 | 15473 | 19060 | 20042 | 7956 | 1628 | 1188 | 1151 |
| 2006 | 12200 | 26690 | 9986 | 11286 | 8395 | 3824 | 1008 | 1281 |
| 2007 | 15181 | 10163 | 19157 | 7078 | 5564 | 3610 | 1218 | 930 |
| 2008 | 6924 | 23230 | 10930 | 14196 | 4977 | 3276 | 3551 | 3118 |
| 2009 | 6607 | 14349 | 13827 | 5817 | 8419 | 2978 | 1505 | 2934 |
| 2010 | 7880 | 8859 | 9174 | 6394 | 2670 | 4762 | 1679 | 2669 |
| 2011 | 10150 | 22799 | 8852 | 3630 | 2860 | 1183 | 1563 | 3869 |
| 2012 | 7029 | 11712 | 15572 | 4016 | 1971 | 1267 | 537 | 2610 |
| 2013 | 4999 | 25516 | 4974 | 7645 | 1886 | 1241 | 616 | 1658 |
| 2014 | 3099 | 12117 | 13380 | 3737 | 4047 | 1036 | 429 | 1388 |
| $2015$ | 6206 | 7392 | 13555 | 8021 | 1844 | 1621 | 715 | 975 |

Table 11.3.8. Saithe in Subareas 4 and 6 and Division 3.a. Dicards numbers (thousands) at age for the age range used in the assessment.

| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 9617 | 3175 | 1141 | 55 | 16 | 7 | 5 | 2 |
| 1968 | 12888 | 4156 | 969 | 143 | 4 | 6 | 3 | 2 |
| 1969 | 16779 | 3461 | 813 | 181 | 57 | 19 | 16 | 2 |
| 1970 | 20671 | 15007 | 1102 | 288 | 38 | 42 | 8 | 3 |
| 1971 | 38741 | 11313 | 2406 | 112 | 42 | 67 | 48 | 9 |
| 1972 | 26680 | 12938 | 1861 | 504 | 36 | 69 | 42 | 18 |
| 1973 | 30152 | 7380 | 1973 | 381 | 83 | 47 | 35 | 26 |
| 1974 | 37148 | 6605 | 1038 | 491 | 144 | 122 | 38 | 22 |
| 1975 | 31626 | 5063 | 758 | 180 | 135 | 220 | 89 | 30 |
| 1976 | 115333 | 10387 | 860 | 148 | 39 | 126 | 92 | 38 |
| 1977 | 15240 | 10760 | 1084 | 130 | 37 | 81 | 97 | 52 |
| 1978 | 19456 | 5338 | 1327 | 79 | 11 | 24 | 20 | 38 |
| 1979 | 9063 | 2783 | 822 | 207 | 28 | 17 | 14 | 23 |
| 1980 | 9709 | 2416 | 662 | 160 | 56 | 27 | 13 | 25 |
| 1981 | 14527 | 1632 | 495 | 87 | 33 | 80 | 17 | 28 |
| 1982 | 17700 | 7945 | 674 | 142 | 21 | 34 | 25 | 14 |
| 1983 | 15673 | 4033 | 1912 | 135 | 48 | 35 | 26 | 15 |
| 1984 | 44915 | 6298 | 952 | 309 | 16 | 29 | 9 | 7 |
| 1985 | 74378 | 10885 | 976 | 113 | 30 | 16 | 10 | 8 |
| 1986 | 30764 | 17857 | 1116 | 128 | 26 | 34 | 11 | 10 |
| 1987 | 17326 | 19080 | 1028 | 75 | 18 | 29 | 23 | 8 |
| 1988 | 18079 | 5169 | 2596 | 91 | 19 | 26 | 20 | 7 |
| 1989 | 12280 | 6020 | 969 | 260 | 15 | 17 | 11 | 5 |
| 1990 | 22647 | 3833 | 832 | 112 | 27 | 18 | 7 | 4 |
| 1991 | 25593 | 5847 | 549 | 85 | 17 | 22 | 9 | 5 |
| 1992 | 10213 | 6580 | 937 | 76 | 15 | 16 | 13 | 5 |
| 1993 | 20990 | 4077 | 871 | 74 | 15 | 36 | 21 | 11 |
| 1994 | 11076 | 7868 | 958 | 102 | 9 | 8 | 11 | 9 |
| 1995 | 14797 | 5096 | 1087 | 90 | 34 | 16 | 14 | 9 |
| 1996 | 6141 | 7607 | 866 | 184 | 31 | 19 | 5 | 5 |
| 1997 | 8344 | 3778 | 2218 | 87 | 26 | 24 | 7 | 3 |
| 1998 | 5751 | 6072 | 1203 | 382 | 22 | 28 | 14 | 3 |
| 1999 | 5233 | 2716 | 1961 | 199 | 99 | 55 | 23 | 7 |
| 2000 | 3920 | 3386 | 657 | 293 | 31 | 76 | 16 | 5 |
| 2001 | 8908 | 3454 | 1648 | 80 | 35 | 42 | 41 | 3 |
| 2002 | 8439 | 5710 | 2451 | 425 | 64 | 324 | 121 | 96 |
| 2003 | 15288 | 7106 | 234 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 5605 | 4407 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 3498 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 5114 | 5282 | 394 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 9433 | 3152 | 1762 | 97 | 0 | 0 | 0 | 0 |
| 2008 | 696 | 7682 | 1610 | 745 | 111 | 9 | 0 | 0 |
| 2009 | 831 | 1158 | 395 | 30 | 93 | 16 | 14 | 11 |
| 2010 | 886 | 390 | 266 | 117 | 1 | 11 | 0 | 38 |
| 2011 | 2636 | 1470 | 129 | 44 | 7 | 25 | 1 | 8 |
| 2012 | 7305 | 1341 | 1377 | 58 | 7 | 1 | 4 | 1 |
| 2013 | 2268 | 4801 | 339 | 224 | 4 | 0 | 0 | 1 |
| 2014 | 955 | 2205 | 1816 | 220 | 77 | 4 | 0 | 1 |
| 2015 | 2163 | 931 | 704 | 232 | 17 | 3 | 0 | 2 |

Table 11.3.9. Saithe in Subareas 4 and 6 and Division 3.a. Catch weight-at-age (kg).

| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 0.898 | 1.339 | 2.094 | 3.183 | 3.753 | 5.316 | 5.891 | 7.719 |
| 1968 | 1.234 | 1.624 | 1.979 | 3.007 | 4.039 | 4.428 | 6.136 | 7.406 |
| 1969 | 0.933 | 1.530 | 2.251 | 2.711 | 3.558 | 4.406 | 5.220 | 6.767 |
| 1970 | 0.908 | 1.416 | 2.049 | 2.716 | 3.599 | 4.463 | 5.687 | 6.845 |
| 1971 | 0.811 | 1.325 | 2.167 | 2.934 | 3.765 | 4.634 | 5.172 | 6.163 |
| 1972 | 0.780 | 1.175 | 1.952 | 2.367 | 3.793 | 4.228 | 4.630 | 6.326 |
| 1973 | 0.792 | 1.382 | 1.633 | 2.569 | 3.356 | 4.684 | 4.814 | 6.445 |
| 1974 | 0.831 | 1.534 | 2.372 | 2.751 | 3.428 | 4.498 | 5.713 | 7.857 |
| 1975 | 0.862 | 1.472 | 2.479 | 3.298 | 3.764 | 4.296 | 5.540 | 7.562 |
| 1976 | 0.678 | 1.287 | 2.250 | 3.068 | 4.034 | 4.383 | 5.112 | 7.147 |
| 1977 | 0.733 | 1.234 | 1.926 | 3.108 | 4.161 | 4.605 | 4.859 | 6.542 |
| 1978 | 0.793 | 1.304 | 2.145 | 3.338 | 4.521 | 4.900 | 5.449 | 7.400 |
| 1979 | 1.069 | 1.595 | 2.228 | 3.093 | 4.049 | 5.274 | 6.308 | 7.955 |
| 1980 | 0.921 | 1.790 | 2.380 | 3.028 | 4.089 | 5.126 | 5.939 | 8.148 |
| 1981 | 0.927 | 1.790 | 2.705 | 3.584 | 4.535 | 5.478 | 6.980 | 8.724 |
| 1982 | 1.048 | 1.548 | 2.518 | 3.218 | 4.206 | 5.125 | 5.905 | 8.823 |
| 1983 | 0.992 | 1.688 | 2.139 | 3.135 | 3.690 | 4.632 | 5.505 | 8.453 |
| 1984 | 0.767 | 1.586 | 2.286 | 2.688 | 3.895 | 4.665 | 6.183 | 8.474 |
| 1985 | 0.640 | 1.244 | 1.941 | 2.769 | 3.406 | 4.950 | 5.865 | 8.854 |
| 1986 | 0.670 | 1.018 | 1.786 | 2.430 | 3.571 | 4.209 | 5.651 | 8.218 |
| 1987 | 0.650 | 0.861 | 1.815 | 3.072 | 4.209 | 5.330 | 6.128 | 8.603 |
| 1988 | 0.751 | 0.964 | 1.379 | 2.789 | 4.023 | 5.254 | 6.322 | 8.649 |
| 1989 | 0.864 | 1.018 | 1.413 | 1.997 | 3.913 | 5.017 | 6.430 | 8.431 |
| 1990 | 0.815 | 1.175 | 1.575 | 2.245 | 3.241 | 4.858 | 6.315 | 8.416 |
| 1991 | 0.764 | 1.138 | 1.744 | 2.363 | 3.165 | 4.222 | 6.066 | 8.191 |
| 1992 | 0.930 | 1.169 | 1.599 | 2.240 | 3.667 | 4.330 | 5.412 | 7.045 |
| 1993 | 0.868 | 1.239 | 1.746 | 2.634 | 3.184 | 3.980 | 5.080 | 6.891 |
| 1994 | 0.911 | 1.100 | 1.594 | 2.432 | 3.617 | 4.787 | 6.548 | 8.326 |
| 1995 | 0.967 | 1.272 | 1.807 | 2.560 | 3.554 | 4.767 | 5.267 | 7.891 |
| 1996 | 0.933 | 1.167 | 1.798 | 2.366 | 2.951 | 4.705 | 6.092 | 8.382 |
| 1997 | 0.873 | 1.125 | 1.445 | 2.585 | 3.555 | 4.525 | 6.158 | 8.866 |
| 1998 | 0.861 | 0.949 | 1.386 | 1.743 | 2.948 | 3.883 | 4.996 | 7.227 |
| 1999 | 0.850 | 1.042 | 1.206 | 1.752 | 2.337 | 3.493 | 4.844 | 6.745 |
| 2000 | 0.992 | 1.107 | 1.532 | 1.683 | 2.593 | 3.084 | 4.773 | 7.461 |
| 2001 | 0.774 | 1.053 | 1.307 | 2.093 | 2.546 | 3.485 | 4.141 | 6.141 |
| 2002 | 0.776 | 1.014 | 1.495 | 1.791 | 2.961 | 3.761 | 4.638 | 5.750 |
| 2003 | 0.636 | 0.889 | 1.167 | 1.810 | 2.368 | 3.176 | 3.768 | 5.065 |
| 2004 | 0.794 | 1.010 | 1.392 | 1.896 | 2.860 | 3.687 | 4.814 | 7.059 |
| 2005 | 0.715 | 1.155 | 1.325 | 1.710 | 2.132 | 3.026 | 3.622 | 5.713 |
| 2006 | 0.904 | 1.012 | 1.489 | 1.906 | 2.424 | 3.058 | 4.318 | 5.734 |
| 2007 | 0.769 | 1.124 | 1.286 | 1.834 | 2.328 | 2.887 | 3.600 | 4.975 |
| 2008 | 0.916 | 1.065 | 1.488 | 1.692 | 2.210 | 2.792 | 3.206 | 4.565 |
| 2009 | 1.033 | 1.333 | 1.672 | 1.994 | 2.566 | 3.086 | 3.651 | 4.790 |
| 2010 | 1.037 | 1.474 | 2.033 | 2.597 | 3.163 | 3.488 | 3.968 | 5.223 |
| 2011 | 0.955 | 1.192 | 1.787 | 2.571 | 3.068 | 3.418 | 3.718 | 4.289 |
| 2012 | 0.910 | 1.287 | 1.383 | 2.196 | 3.221 | 3.536 | 4.181 | 4.482 |
| 2013 | 0.878 | 1.132 | 1.586 | 1.957 | 3.076 | 3.841 | 4.541 | 5.648 |
| 2014 | 1.091 | 1.265 | 1.568 | 2.334 | 2.607 | 4.010 | 5.530 | 6.679 |
| 2015 | 0.951 | 1.253 | 1.621 | 2.180 | 3.037 | 3.793 | 4.228 | 7.285 |

Table 11.3.10. Saithe in Subareas 4 and 6 and Division 3.a. Landings weight-at-age (kg).

| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 0.9305 | 1.3620 | 2.1035 | 3.1858 | 3.7541 | 5.3162 | 5.8905 | 7.7190 |
| 1968 | 1.2784 | 1.6521 | 1.9886 | 3.0093 | 4.0404 | 4.4278 | 6.1355 | 7.4055 |
| 1969 | 0.9663 | 1.5568 | 2.2614 | 2.7133 | 3.5588 | 4.4063 | 5.2203 | 6.7675 |
| 1970 | 0.9414 | 1.4408 | 2.0587 | 2.7180 | 3.5995 | 4.4632 | 5.6871 | 6.8452 |
| 1971 | 0.8399 | 1.3480 | 2.1775 | 2.9360 | 3.7657 | 4.6339 | 5.1725 | 6.1630 |
| 1972 | 0.8082 | 1.1958 | 1.9610 | 2.3687 | 3.7941 | 4.2276 | 4.6304 | 6.3263 |
| 1973 | 0.8212 | 1.4061 | 1.6410 | 2.5709 | 3.3571 | 4.6844 | 4.8138 | 6.4449 |
| 1974 | 0.8608 | 1.5606 | 2.3834 | 2.7527 | 3.4286 | 4.4977 | 5.7128 | 7.8570 |
| 1975 | 0.8928 | 1.4977 | 2.4904 | 3.3002 | 3.7647 | 4.2957 | 5.5396 | 7.5620 |
| 1976 | 0.7024 | 1.3092 | 2.2604 | 3.0706 | 4.0347 | 4.3833 | 5.1117 | 7.1470 |
| 1977 | 0.7598 | 1.2560 | 1.9348 | 3.1107 | 4.1618 | 4.6045 | 4.8589 | 6.5419 |
| 1978 | 0.8215 | 1.3267 | 2.1545 | 3.3401 | 4.5221 | 4.9005 | 5.4494 | 7.4000 |
| 1979 | 1.1072 | 1.6228 | 2.2381 | 3.0950 | 4.0504 | 5.2742 | 6.3077 | 7.9551 |
| 1980 | 0.9546 | 1.8212 | 2.3911 | 3.0300 | 4.0895 | 5.1262 | 5.9393 | 8.1476 |
| 1981 | 0.9608 | 1.8211 | 2.7175 | 3.5868 | 4.5360 | 5.4776 | 6.9804 | 8.7237 |
| 1982 | 1.0857 | 1.5746 | 2.5293 | 3.2202 | 4.2069 | 5.1251 | 5.9049 | 8.8232 |
| 1983 | 1.0276 | 1.7178 | 2.1493 | 3.1377 | 3.6906 | 4.6317 | 5.5053 | 8.4529 |
| 1984 | 0.7948 | 1.6139 | 2.2966 | 2.6899 | 3.8959 | 4.6647 | 6.1830 | 8.4735 |
| 1985 | 0.6632 | 1.2654 | 1.9505 | 2.7715 | 3.4067 | 4.9499 | 5.8649 | 8.8543 |
| 1986 | 0.6943 | 1.0353 | 1.7944 | 2.4316 | 3.5717 | 4.2094 | 5.6506 | 8.2184 |
| 1987 | 0.6739 | 0.8763 | 1.8236 | 3.0747 | 4.2098 | 5.3300 | 6.1284 | 8.6026 |
| 1988 | 0.7787 | 0.9810 | 1.3859 | 2.7907 | 4.0238 | 5.2544 | 6.3221 | 8.6489 |
| 1989 | 0.8954 | 1.0362 | 1.4196 | 1.9984 | 3.9139 | 5.0175 | 6.4298 | 8.4308 |
| 1990 | 0.8441 | 1.1958 | 1.5828 | 2.2472 | 3.2419 | 4.8583 | 6.3149 | 8.4162 |
| 1991 | 0.7913 | 1.1579 | 1.7523 | 2.3646 | 3.1653 | 4.2221 | 6.0661 | 8.1914 |
| 1992 | 0.9641 | 1.1893 | 1.6066 | 2.2417 | 3.6677 | 4.3296 | 5.4125 | 7.0455 |
| 1993 | 0.8994 | 1.2603 | 1.7544 | 2.6363 | 3.1851 | 3.9798 | 5.0802 | 6.8909 |
| 1994 | 0.9439 | 1.1188 | 1.6010 | 2.4337 | 3.6175 | 4.7869 | 6.5479 | 8.3256 |
| 1995 | 1.0022 | 1.2937 | 1.8159 | 2.5619 | 3.5549 | 4.7670 | 5.2674 | 7.8907 |
| 1996 | 0.9668 | 1.1873 | 1.8068 | 2.3678 | 2.9518 | 4.7053 | 6.0922 | 8.3821 |
| 1997 | 0.9047 | 1.1448 | 1.4522 | 2.5867 | 3.5556 | 4.5251 | 6.1575 | 8.8663 |
| 1998 | 0.8917 | 0.9660 | 1.3925 | 1.7440 | 2.9486 | 3.8829 | 4.9955 | 7.2273 |
| 1999 | 0.8808 | 1.0605 | 1.2112 | 1.7537 | 2.3374 | 3.4934 | 4.8438 | 6.7452 |
| 2000 | 1.0274 | 1.1266 | 1.5389 | 1.6843 | 2.5936 | 3.0842 | 4.7733 | 7.4615 |
| 2001 | 0.8023 | 1.0717 | 1.3130 | 2.0950 | 2.5461 | 3.4848 | 4.1410 | 6.1410 |
| 2002 | 0.9233 | 1.0348 | 1.4777 | 1.7691 | 2.9469 | 3.4261 | 4.4066 | 5.6741 |
| 2003 | 0.8327 | 0.9801 | 1.1732 | 1.8103 | 2.3683 | 3.1761 | 3.7684 | 5.0647 |
| 2004 | 0.9182 | 1.0839 | 1.3915 | 1.8959 | 2.8599 | 3.6872 | 4.8135 | 7.0589 |
| 2005 | 0.9211 | 1.1553 | 1.3252 | 1.7095 | 2.1315 | 3.0262 | 3.6217 | 5.7133 |
| 2006 | 0.9445 | 1.0687 | 1.5137 | 1.9060 | 2.4242 | 3.0581 | 4.3175 | 5.7338 |
| 2007 | 0.8369 | 1.1427 | 1.3168 | 1.8401 | 2.3283 | 2.8874 | 3.6002 | 4.9754 |
| 2008 | 0.9444 | 1.1925 | 1.5650 | 1.7199 | 2.2264 | 2.7948 | 3.2060 | 4.5654 |
| 2009 | 1.0357 | 1.3396 | 1.6638 | 1.9920 | 2.5627 | 3.0845 | 3.6483 | 4.7929 |
| 2010 | 1.0359 | 1.4786 | 2.0343 | 2.5974 | 3.1636 | 3.4884 | 3.9677 | 5.1988 |
| 2011 | 1.0072 | 1.2065 | 1.7828 | 2.5727 | 3.0682 | 3.4043 | 3.7174 | 4.2837 |
| 2012 | 1.0151 | 1.3207 | 1.4080 | 2.2014 | 3.2230 | 3.5363 | 4.1772 | 4.4822 |
| 2013 | 0.8978 | 1.1563 | 1.6141 | 1.9761 | 3.0780 | 3.8405 | 4.5407 | 5.6475 |
| 2014 | 1.1264 | 1.3004 | 1.6074 | 2.3842 | 2.6172 | 4.0126 | 5.5301 | 6.6790 |
| 2015 | 0.9770 | 1.2441 | 1.6253 | 2.1899 | 3.0431 | 3.7957 | 4.2282 | 7.2869 |

Table 11.3.11. Saithe in Subareas 4 and 6 and Division 3.a. Discards weight-at-age (kg).

| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 0.748 | 1.076 | 1.818 | 2.972 | 3.590 | 5.316 | 5.891 | 7.719 |
| 1968 | 1.028 | 1.306 | 1.719 | 2.808 | 3.864 | 4.428 | 6.136 | 7.406 |
| 1969 | 0.777 | 1.230 | 1.955 | 2.531 | 3.403 | 4.406 | 5.220 | 6.767 |
| 1970 | 0.757 | 1.139 | 1.780 | 2.536 | 3.442 | 4.463 | 5.687 | 6.845 |
| 1971 | 0.676 | 1.065 | 1.882 | 2.739 | 3.601 | 4.634 | 5.172 | 6.163 |
| 1972 | 0.650 | 0.945 | 1.695 | 2.210 | 3.628 | 4.228 | 4.630 | 6.326 |
| 1973 | 0.660 | 1.111 | 1.419 | 2.399 | 3.210 | 4.684 | 4.814 | 6.445 |
| 1974 | 0.692 | 1.233 | 2.060 | 2.568 | 3.279 | 4.498 | 5.713 | 7.857 |
| 1975 | 0.718 | 1.184 | 2.153 | 3.079 | 3.600 | 4.296 | 5.540 | 7.562 |
| 1976 | 0.565 | 1.035 | 1.954 | 2.865 | 3.858 | 4.383 | 5.112 | 7.147 |
| 1977 | 0.611 | 0.993 | 1.673 | 2.902 | 3.980 | 4.605 | 4.859 | 6.542 |
| 1978 | 0.661 | 1.049 | 1.862 | 3.116 | 4.325 | 4.900 | 5.449 | 7.400 |
| 1979 | 0.890 | 1.283 | 1.935 | 2.888 | 3.873 | 5.274 | 6.308 | 7.955 |
| 1980 | 0.768 | 1.439 | 2.067 | 2.827 | 3.911 | 5.126 | 5.939 | 8.148 |
| 1981 | 0.773 | 1.439 | 2.349 | 3.346 | 4.338 | 5.478 | 6.980 | 8.724 |
| 1982 | 0.873 | 1.245 | 2.186 | 3.004 | 4.023 | 5.125 | 5.905 | 8.823 |
| 1983 | 0.826 | 1.358 | 1.858 | 2.927 | 3.529 | 4.632 | 5.505 | 8.453 |
| 1984 | 0.639 | 1.276 | 1.985 | 2.510 | 3.726 | 4.665 | 6.183 | 8.474 |
| 1985 | 0.533 | 1.000 | 1.686 | 2.586 | 3.258 | 4.950 | 5.865 | 8.854 |
| 1986 | 0.558 | 0.818 | 1.551 | 2.269 | 3.416 | 4.209 | 5.651 | 8.218 |
| 1987 | 0.542 | 0.693 | 1.576 | 2.869 | 4.026 | 5.330 | 6.128 | 8.603 |
| 1988 | 0.626 | 0.775 | 1.198 | 2.604 | 3.848 | 5.254 | 6.322 | 8.649 |
| 1989 | 0.720 | 0.819 | 1.227 | 1.865 | 3.743 | 5.017 | 6.430 | 8.431 |
| 1990 | 0.679 | 0.945 | 1.368 | 2.097 | 3.100 | 4.858 | 6.315 | 8.416 |
| 1991 | 0.636 | 0.915 | 1.515 | 2.206 | 3.027 | 4.222 | 6.066 | 8.191 |
| 1992 | 0.775 | 0.940 | 1.389 | 2.092 | 3.508 | 4.330 | 5.412 | 7.045 |
| 1993 | 0.723 | 0.996 | 1.517 | 2.460 | 3.046 | 3.980 | 5.080 | 6.891 |
| 1994 | 0.759 | 0.884 | 1.384 | 2.271 | 3.459 | 4.787 | 6.548 | 8.326 |
| 1995 | 0.806 | 1.023 | 1.570 | 2.390 | 3.400 | 4.767 | 5.267 | 7.891 |
| 1996 | 0.778 | 0.938 | 1.562 | 2.209 | 2.823 | 4.705 | 6.092 | 8.382 |
| 1997 | 0.728 | 0.905 | 1.255 | 2.413 | 3.400 | 4.525 | 6.158 | 8.866 |
| 1998 | 0.717 | 0.764 | 1.204 | 1.627 | 2.820 | 3.883 | 4.996 | 7.227 |
| 1999 | 0.708 | 0.838 | 1.047 | 1.636 | 2.235 | 3.493 | 4.844 | 6.745 |
| 2000 | 0.826 | 0.890 | 1.330 | 1.571 | 2.480 | 3.084 | 4.773 | 7.461 |
| 2001 | 0.645 | 0.847 | 1.135 | 1.955 | 2.435 | 3.485 | 4.141 | 6.141 |
| 2002 | 0.616 | 0.896 | 1.580 | 2.483 | 3.469 | 6.058 | 6.935 | 6.927 |
| 2003 | 0.469 | 0.571 | 0.641 | 1.689 | 2.265 | 3.176 | 3.768 | 5.065 |
| 2004 | 0.617 | 0.676 | 1.203 | 1.769 | 2.735 | 3.687 | 4.814 | 7.059 |
| 2005 | 0.741 | 0.913 | 1.146 | 1.595 | 2.038 | 3.026 | 3.622 | 5.713 |
| 2006 | 0.808 | 0.724 | 0.859 | 1.778 | 2.318 | 3.058 | 4.318 | 5.734 |
| 2007 | 0.660 | 1.062 | 0.949 | 1.365 | 2.227 | 2.887 | 3.600 | 4.975 |
| 2008 | 0.633 | 0.680 | 0.967 | 1.161 | 1.495 | 1.820 | 3.206 | 2.797 |
| 2009 | 1.010 | 1.253 | 1.946 | 2.403 | 2.838 | 3.388 | 3.934 | 3.911 |
| 2010 | 1.046 | 1.374 | 1.987 | 2.561 | 3.025 | 3.351 | 3.968 | 6.895 |
| 2011 | 0.756 | 0.971 | 2.054 | 2.445 | 3.170 | 4.072 | 4.369 | 6.618 |
| 2012 | 0.808 | 0.997 | 1.101 | 1.831 | 2.675 | 3.411 | 4.804 | 5.313 |
| 2013 | 0.835 | 1.003 | 1.180 | 1.300 | 2.298 | 3.841 | 4.541 | 5.861 |
| 2014 | 0.977 | 1.072 | 1.274 | 1.487 | 2.077 | 3.223 | 5.530 | 7.568 |
| 2015 | 0.877 | 1.326 | 1.531 | 1.848 | 2.410 | 2.184 | 4.228 | 5.911 |

Table 11.3.12. Saithe in Subareas 4 and 6 and Division 3.a. Data available for calibration of the assessment. Indices include one commercial standardized CPUE indices (year effects), tuned to the exploitable biomass within SAM, and one research survey.

| IBTS Q3 (DATRAS STANDARD INDEX) |  |  |  |  |  |  | CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 3 | 4 | 5 | 6 | 7 | 8 |  |
| 1992 | 1.077 | 2.76 | 0.516 | 0.098 | 0.057 | 0.05 |  |
| 1993 | 7.965 | 2.781 | 1.129 | 0.197 | 0.011 | 0.04 |  |
| 1994 | 1.117 | 1.615 | 0.893 | 0.609 | 0.091 | 0.04 |  |
| 1995 | 13.959 | 2.501 | 1.559 | 0.533 | 0.172 | 0.049 |  |
| 1996 | 3.825 | 6.533 | 1.112 | 0.971 | 0.212 | 0.069 |  |
| 1997 | 3.756 | 3.351 | 7.461 | 0.698 | 0.534 | 0.181 |  |
| 1998 | 1.181 | 4.134 | 1.351 | 1.58 | 0.149 | 0.179 |  |
| 1999 | 2.086 | 1.907 | 3.155 | 0.619 | 0.632 | 0.074 |  |
| 2000 | 3.479 | 8.836 | 1.081 | 0.868 | 0.114 | 0.152 | 0.217 |
| 2001 | 21.614 | 6.206 | 3.959 | 0.357 | 0.446 | 0.114 | 0.3368 |
| 2002 | 10.748 | 18.974 | 1.327 | 1.09 | 0.162 | 0.264 | 0.1296 |
| 2003 | 19.272 | 23.802 | 13.402 | 0.393 | 0.439 | 0.168 | 0.05785 |
| 2004 | 4.93 | 6.727 | 3.237 | 0.921 | 0.064 | 0.085 | 0.2981 |
| 2005 | 8.916 | 7.512 | 4.428 | 1.914 | 1.082 | 0.104 | 0.391 |
| 2006 | 10.553 | 29.579 | 2.835 | 1.177 | 0.445 | 0.242 | 0.4279 |
| 2007 | 34.006 | 5.578 | 11.7 | 1.016 | 0.743 | 0.358 | 0.2428 |
| 2008 | 3.312 | 5.584 | 0.907 | 1.997 | 0.254 | 0.254 | 0.4292 |
| 2009 | 1.346 | 1.703 | 0.568 | 0.101 | 0.229 | 0.2 | 0.1744 |
| 2010 | 1.361 | 0.964 | 0.471 | 0.205 | 0.045 | 0.166 | 0.1069 |
| 2011 | 4.52 | 8.451 | 1.059 | 1.114 | 0.426 | 0.08 | 0.0939 |
| 2012 | 11.134 | 2.497 | 2.968 | 0.503 | 0.483 | 0.344 | -0.0428 |
| 2013 | 14.701 | 16.279 | 1.83 | 1.858 | 0.308 | 0.146 | 0.05277 |
| 2014 | 1.649 | 3.923 | 2.822 | 0.481 | 0.52 | 0.114 | -0.01323 |
| 2015 | 11.001 | 5.613 | 4.611 | 1.581 | 0.289 | 0.285 | 0.1663 |

Table 11.4.1. Saithe in Subareas 4 and 6 and Division 3.a. Model configuration for both the final and the exploratory SAM assessment.

Min Age: 3
Max Age: 10
Max Age considered a plus group (Yes)
The following matrix describes the coupling of fishing mortality STATES, where rows represent fleets and columns represent ages:

| 1 | 2 | 3 | 4 | 5 | 6 | 7 | 7 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Use correlated random walks for the fishing mortalities: (2=AR1)
2
Coupling of catchability PARAMETERS

| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2 | 3 | 4 | 5 | 6 | 0 | 0 |
| Coupling of power law model EXPONENTS (if used) |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Coupling of fishing mortality RW VARIANCES

| 1 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Coupling of $\log \mathrm{N}$ |  |  |  |  |  |  |  |
| RW | VARIANCES |  |  |  |  |  |  |


| 1 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Coupling of OBSERVATION VARIANCES

| 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 2 | 2 | 2 | 2 | 2 | 0 | 0 |

Stock recruitment model code (random walk)
Years in which catch data are to be scaled by an estimated parameter

## 0

Fbar range: 4 to 7
Observation correlation coupling ( $0=$ uncorrelated). Rows represent fleets, columns represent adjacent age groups, i.e. the first column is the correlation between the first and 2nd age group. An NA in all non-empty age groups for a fleet specifies unstructured correlation. NA's and positive numbers cannot be mixed within fleets.

| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| NA | NA | NA | NA | NA | $\mathbf{0}$ | $\mathbf{0}$ |

Table 11.4.2. Saithe in Subareas 4 and 6 and Division 3.a. Fishing mortalities at age for the exploratory assessment model: cpue + DATRAS Q3, sw=cw for ages 7+. Fs for ages 9 and 10+ were coupled in the model configuration.

| Year\Age | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 0.268 | 0.385 | 0.355 | 0.350 | 0.311 | 0.281 | 0.310 |
| 1968 | 0.238 | 0.344 | 0.301 | 0.284 | 0.245 | 0.221 | 0.247 |
| 1969 | 0.257 | 0.371 | 0.324 | 0.311 | 0.276 | 0.252 | 0.273 |
| 1970 | 0.305 | 0.418 | 0.351 | 0.326 | 0.282 | 0.253 | 0.266 |
| 1971 | 0.370 | 0.467 | 0.377 | 0.345 | 0.308 | 0.284 | 0.295 |
| 1972 | 0.445 | 0.521 | 0.405 | 0.368 | 0.331 | 0.306 | 0.310 |
| 1973 | 0.520 | 0.572 | 0.429 | 0.381 | 0.345 | 0.319 | 0.317 |
| 1974 | 0.631 | 0.662 | 0.497 | 0.436 | 0.398 | 0.364 | 0.350 |
| 1975 | 0.655 | 0.695 | 0.537 | 0.476 | 0.442 | 0.409 | 0.385 |
| 1976 | 0.743 | 0.775 | 0.609 | 0.531 | 0.486 | 0.445 | 0.409 |
| 1977 | 0.633 | 0.714 | 0.598 | 0.541 | 0.510 | 0.473 | 0.430 |
| 1978 | 0.501 | 0.586 | 0.492 | 0.439 | 0.416 | 0.388 | 0.355 |
| 1979 | 0.419 | 0.523 | 0.459 | 0.422 | 0.408 | 0.380 | 0.346 |
| 1980 | 0.405 | 0.522 | 0.479 | 0.453 | 0.447 | 0.423 | 0.386 |
| 1981 | 0.366 | 0.500 | 0.473 | 0.459 | 0.464 | 0.452 | 0.416 |
| 1982 | 0.432 | 0.586 | 0.553 | 0.521 | 0.508 | 0.481 | 0.436 |
| 1983 | 0.515 | 0.705 | 0.674 | 0.626 | 0.597 | 0.554 | 0.493 |
| 1984 | 0.586 | 0.793 | 0.725 | 0.628 | 0.561 | 0.504 | 0.444 |
| 1985 | 0.626 | 0.869 | 0.772 | 0.625 | 0.541 | 0.483 | 0.436 |
| 1986 | 0.592 | 0.897 | 0.821 | 0.654 | 0.565 | 0.511 | 0.478 |
| 1987 | 0.540 | 0.845 | 0.794 | 0.632 | 0.553 | 0.509 | 0.490 |
| 1988 | 0.528 | 0.832 | 0.804 | 0.647 | 0.567 | 0.521 | 0.503 |
| 1989 | 0.517 | 0.813 | 0.783 | 0.628 | 0.538 | 0.483 | 0.464 |
| 1990 | 0.500 | 0.784 | 0.749 | 0.591 | 0.501 | 0.439 | 0.422 |
| 1991 | 0.463 | 0.745 | 0.719 | 0.565 | 0.479 | 0.418 | 0.408 |
| 1992 | 0.410 | 0.694 | 0.695 | 0.559 | 0.482 | 0.419 | 0.413 |
| 1993 | 0.389 | 0.681 | 0.710 | 0.603 | 0.559 | 0.501 | 0.499 |
| 1994 | 0.319 | 0.595 | 0.630 | 0.539 | 0.512 | 0.466 | 0.472 |
| 1995 | 0.275 | 0.553 | 0.618 | 0.559 | 0.561 | 0.529 | 0.536 |
| 1996 | 0.218 | 0.466 | 0.545 | 0.507 | 0.506 | 0.481 | 0.483 |
| 1997 | 0.184 | 0.403 | 0.476 | 0.446 | 0.436 | 0.419 | 0.419 |
| 1998 | 0.184 | 0.403 | 0.485 | 0.461 | 0.440 | 0.423 | 0.419 |
| 1999 | 0.182 | 0.410 | 0.512 | 0.506 | 0.484 | 0.471 | 0.461 |
| 2000 | 0.155 | 0.360 | 0.450 | 0.444 | 0.407 | 0.388 | 0.377 |
| 2001 | 0.147 | 0.341 | 0.418 | 0.410 | 0.364 | 0.342 | 0.334 |
| 2002 | 0.143 | 0.335 | 0.420 | 0.438 | 0.400 | 0.388 | 0.401 |
| 2003 | 0.149 | 0.337 | 0.416 | 0.456 | 0.426 | 0.415 | 0.435 |
| 2004 | 0.133 | 0.309 | 0.372 | 0.406 | 0.374 | 0.360 | 0.365 |
| 2005 | 0.135 | 0.316 | 0.381 | 0.415 | 0.380 | 0.360 | 0.354 |
| 2006 | 0.150 | 0.341 | 0.400 | 0.425 | 0.388 | 0.366 | 0.357 |
| 2007 | 0.143 | 0.335 | 0.389 | 0.402 | 0.364 | 0.341 | 0.332 |
| 2008 | 0.157 | 0.383 | 0.459 | 0.470 | 0.426 | 0.399 | 0.385 |
| 2009 | 0.152 | 0.382 | 0.467 | 0.480 | 0.438 | 0.413 | 0.397 |
| 2010 | 0.137 | 0.359 | 0.445 | 0.459 | 0.429 | 0.413 | 0.398 |
| 2011 | 0.143 | 0.372 | 0.454 | 0.455 | 0.418 | 0.402 | 0.385 |
| 2012 | 0.123 | 0.339 | 0.414 | 0.417 | 0.381 | 0.367 | 0.355 |
| 2013 | 0.103 | 0.299 | 0.367 | 0.375 | 0.347 | 0.338 | 0.330 |
| 2014 | 0.089 | 0.265 | 0.332 | 0.344 | 0.321 | 0.317 | 0.319 |
| 2015 | 0.088 | 0.263 | 0.328 | 0.335 | 0.309 | 0.301 | 0.304 |

Table 11.4.3. Saithe in Subareas 4 and 6 and Division 3.a. Estimated population numbers-at-age for the exploratory assessment model: cpue + DATRAS Q3, sw=cw for ages 7+.

| Year \AGe | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 140963 | 81613 | 57347 | 7177 | 4923 | 1158 | 759 | 694 |
| 1968 | 160362 | 91876 | 50370 | 31878 | 3747 | 2547 | 665 | 793 |
| 1969 | 283740 | 90417 | 54375 | 31055 | 20519 | 2842 | 1966 | 835 |
| 1970 | 293388 | 215022 | 49090 | 35300 | 18682 | 11735 | 1805 | 1646 |
| 1971 | 354167 | 191110 | 118659 | 24662 | 19415 | 11927 | 7836 | 2520 |
| 1972 | 224084 | 209036 | 102790 | 67090 | 14503 | 11360 | 7303 | 6511 |
| 1973 | 201100 | 111374 | 105146 | 63086 | 35662 | 8668 | 6336 | 8605 |
| 1974 | 199243 | 90766 | 48379 | 62491 | 41760 | 20472 | 5383 | 8537 |
| 1975 | 234909 | 76841 | 35525 | 24203 | 35998 | 24968 | 11932 | 8471 |
| 1976 | 403670 | 102762 | 29696 | 17386 | 12851 | 19027 | 13191 | 11534 |
| 1977 | 149927 | 148828 | 35701 | 12435 | 8653 | 7165 | 10688 | 13908 |
| 1978 | 120345 | 72388 | 58174 | 14262 | 5124 | 4006 | 3384 | 13088 |
| 1979 | 87539 | 53928 | 34763 | 29293 | 7801 | 2815 | 2207 | 9519 |
| 1980 | 85562 | 47065 | 25688 | 18703 | 16078 | 4041 | 1667 | 7666 |
| 1981 | 162508 | 41887 | 24867 | 12280 | 9624 | 8280 | 2144 | 5881 |
| 1982 | 140911 | 107955 | 22924 | 15011 | 6271 | 4820 | 3773 | 4076 |
| 1983 | 148652 | 69639 | 54668 | 11311 | 8240 | 3131 | 2523 | 3824 |
| 1984 | 255031 | 76154 | 29993 | 23769 | 4726 | 3472 | 1333 | 2793 |
| 1985 | 354922 | 108294 | 29490 | 12780 | 9453 | 2220 | 1591 | 2299 |
| 1986 | 289819 | 141896 | 32321 | 11783 | 6357 | 4473 | 1190 | 2257 |
| 1987 | 149255 | 163455 | 36430 | 10225 | 5136 | 3280 | 2284 | 1798 |
| 1988 | 137966 | 71459 | 61357 | 11450 | 4549 | 2588 | 1737 | 1932 |
| 1989 | 102533 | 69267 | 27683 | 21756 | 4710 | 2088 | 1241 | 1651 |
| 1990 | 151040 | 48014 | 25676 | 11125 | 8371 | 2305 | 1025 | 1408 |
| 1991 | 174004 | 71654 | 17403 | 10254 | 5255 | 3794 | 1234 | 1389 |
| 1992 | 104092 | 88513 | 26170 | 6805 | 5173 | 2850 | 2056 | 1478 |
| 1993 | 175788 | 59386 | 33946 | 9200 | 2916 | 3126 | 1797 | 2250 |
| 1994 | 117944 | 97068 | 28730 | 13421 | 3471 | 1415 | 1469 | 2149 |
| 1995 | 215510 | 66620 | 41962 | 13047 | 6409 | 1628 | 921 | 1926 |
| 1996 | 119756 | 148954 | 29830 | 19658 | 7027 | 2509 | 714 | 1347 |
| 1997 | 148944 | 79834 | 90061 | 13274 | 9296 | 3443 | 1126 | 982 |
| 1998 | 87526 | 120579 | 45524 | 48976 | 7257 | 4650 | 1879 | 1057 |
| 1999 | 111322 | 55243 | 73734 | 22718 | 26719 | 4238 | 2396 | 1626 |
| 2000 | 97469 | 92533 | 29131 | 36771 | 11085 | 12764 | 2024 | 1754 |
| 2001 | 206934 | 67506 | 63561 | 14255 | 17763 | 6384 | 6987 | 1640 |
| 2002 | 163709 | 146677 | 35470 | 34643 | 8349 | 9787 | 3913 | 5082 |
| 2003 | 166719 | 122905 | 85574 | 17033 | 17451 | 5394 | 5375 | 5071 |
| 2004 | 117472 | 111331 | 77223 | 50204 | 8253 | 8455 | 3389 | 4873 |
| 2005 | 143050 | 74983 | 66865 | 49658 | 28050 | 5048 | 4602 | 4481 |
| 2006 | 101651 | 123281 | 41985 | 37019 | 26768 | 14191 | 3093 | 4943 |
| 2007 | 154444 | 55618 | 77627 | 24460 | 19767 | 14839 | 7168 | 4395 |
| 2008 | 73363 | 96788 | 30967 | 47554 | 15269 | 11026 | 9835 | 8026 |
| 2009 | 58434 | 51655 | 43768 | 14616 | 24627 | 9469 | 5785 | 10120 |
| 2010 | 89669 | 37942 | 28176 | 20182 | 7126 | 12960 | 5475 | 9362 |
| 2011 | 82872 | 78852 | 22687 | 14450 | 10156 | 3689 | 6496 | 9785 |
| 2012 | 141079 | 48363 | 48369 | 12264 | 7600 | 5065 | 2011 | 9273 |
| 2013 | 97500 | 107400 | 23753 | 27583 | 7048 | 4113 | 2707 | 6485 |
| 2014 | 58354 | 72343 | 59010 | 13537 | 15405 | 4165 | 2127 | 5301 |
| 2015 | 102967 | 42975 | 49233 | 33743 | 7913 | 8428 | 2648 | 4327 |

Table 11.4.4. Saithe in Subareas 4 and 6 and Division 3.a. Fishing mortalities at age for the final assessment model: cpue + DATRAS Q3, $s w=c w$ for all ages. Fs for ages 9 and $10+$ were coupled in the model configuration.

| Year $\backslash$ Age | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 0.267 | 0.384 | 0.356 | 0.352 | 0.312 | 0.281 | 0.312 |
| 1968 | 0.238 | 0.343 | 0.301 | 0.283 | 0.244 | 0.220 | 0.247 |
| 1969 | 0.256 | 0.369 | 0.324 | 0.311 | 0.276 | 0.252 | 0.274 |
| 1970 | 0.305 | 0.418 | 0.351 | 0.326 | 0.282 | 0.252 | 0.265 |
| 1971 | 0.371 | 0.467 | 0.377 | 0.345 | 0.307 | 0.283 | 0.295 |
| 1972 | 0.446 | 0.521 | 0.404 | 0.367 | 0.330 | 0.306 | 0.310 |
| 1973 | 0.522 | 0.572 | 0.428 | 0.379 | 0.344 | 0.318 | 0.316 |
| 1974 | 0.633 | 0.663 | 0.495 | 0.435 | 0.397 | 0.364 | 0.349 |
| 1975 | 0.655 | 0.696 | 0.535 | 0.475 | 0.442 | 0.409 | 0.385 |
| 1976 | 0.746 | 0.778 | 0.609 | 0.532 | 0.486 | 0.445 | 0.408 |
| 1977 | 0.633 | 0.715 | 0.599 | 0.542 | 0.512 | 0.476 | 0.431 |
| 1978 | 0.501 | 0.585 | 0.490 | 0.438 | 0.415 | 0.388 | 0.353 |
| 1979 | 0.420 | 0.521 | 0.457 | 0.421 | 0.408 | 0.380 | 0.345 |
| 1980 | 0.405 | 0.520 | 0.477 | 0.453 | 0.448 | 0.424 | 0.387 |
| 1981 | 0.364 | 0.496 | 0.470 | 0.458 | 0.466 | 0.454 | 0.418 |
| 1982 | 0.431 | 0.583 | 0.552 | 0.521 | 0.510 | 0.482 | 0.436 |
| 1983 | 0.514 | 0.704 | 0.675 | 0.629 | 0.601 | 0.558 | 0.495 |
| 1984 | 0.586 | 0.794 | 0.726 | 0.628 | 0.561 | 0.504 | 0.443 |
| 1985 | 0.627 | 0.873 | 0.773 | 0.624 | 0.539 | 0.481 | 0.435 |
| 1986 | 0.591 | 0.902 | 0.823 | 0.654 | 0.564 | 0.511 | 0.478 |
| 1987 | 0.538 | 0.848 | 0.795 | 0.631 | 0.551 | 0.509 | 0.490 |
| 1988 | 0.527 | 0.834 | 0.805 | 0.646 | 0.566 | 0.521 | 0.504 |
| 1989 | 0.516 | 0.815 | 0.784 | 0.628 | 0.537 | 0.482 | 0.464 |
| 1990 | 0.500 | 0.787 | 0.750 | 0.590 | 0.499 | 0.437 | 0.420 |
| 1991 | 0.464 | 0.747 | 0.720 | 0.563 | 0.478 | 0.416 | 0.406 |
| 1992 | 0.410 | 0.696 | 0.696 | 0.558 | 0.481 | 0.417 | 0.411 |
| 1993 | 0.389 | 0.683 | 0.712 | 0.603 | 0.560 | 0.501 | 0.501 |
| 1994 | 0.318 | 0.596 | 0.630 | 0.537 | 0.511 | 0.464 | 0.472 |
| 1995 | 0.274 | 0.553 | 0.620 | 0.559 | 0.564 | 0.531 | 0.540 |
| 1996 | 0.218 | 0.465 | 0.546 | 0.507 | 0.508 | 0.483 | 0.486 |
| 1997 | 0.184 | 0.402 | 0.476 | 0.445 | 0.437 | 0.419 | 0.420 |
| 1998 | 0.184 | 0.403 | 0.486 | 0.461 | 0.441 | 0.424 | 0.419 |
| 1999 | 0.182 | 0.411 | 0.515 | 0.509 | 0.487 | 0.474 | 0.464 |
| 2000 | 0.157 | 0.362 | 0.453 | 0.447 | 0.408 | 0.388 | 0.377 |
| 2001 | 0.149 | 0.344 | 0.421 | 0.412 | 0.365 | 0.343 | 0.335 |
| 2002 | 0.144 | 0.334 | 0.420 | 0.440 | 0.404 | 0.394 | 0.412 |
| 2003 | 0.151 | 0.339 | 0.418 | 0.461 | 0.433 | 0.423 | 0.449 |
| 2004 | 0.131 | 0.302 | 0.364 | 0.404 | 0.377 | 0.368 | 0.382 |
| 2005 | 0.135 | 0.315 | 0.379 | 0.416 | 0.381 | 0.361 | 0.358 |
| 2006 | 0.151 | 0.340 | 0.397 | 0.425 | 0.388 | 0.365 | 0.359 |
| 2007 | 0.141 | 0.331 | 0.383 | 0.399 | 0.361 | 0.338 | 0.331 |
| 2008 | 0.159 | 0.386 | 0.461 | 0.473 | 0.425 | 0.394 | 0.377 |
| 2009 | 0.151 | 0.381 | 0.464 | 0.481 | 0.437 | 0.410 | 0.393 |
| 2010 | 0.137 | 0.360 | 0.445 | 0.462 | 0.429 | 0.411 | 0.393 |
| 2011 | 0.143 | 0.373 | 0.454 | 0.457 | 0.419 | 0.403 | 0.385 |
| 2012 | 0.121 | 0.335 | 0.410 | 0.416 | 0.381 | 0.371 | 0.361 |
| 2013 | 0.100 | 0.292 | 0.360 | 0.372 | 0.345 | 0.340 | 0.334 |
| 2014 | 0.086 | 0.257 | 0.323 | 0.338 | 0.316 | 0.314 | 0.317 |
| $2015$ | 0.086 | 0.255 | 0.319 | 0.328 | 0.301 | 0.293 | 0.295 |

Table 11.4.5. Saithe in Subareas 4 and 6 and Division 3.a. Estimated population numbers-at-age for the final assessment model: cpue + DATRAS Q3, $\mathrm{sw}=\mathrm{cw}$ for all ages.

| Year \AGE | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 141049 | 81578 | 57358 | 7176 | 4929 | 1159 | 757 | 692 |
| 1968 | 160710 | 92026 | 50466 | 31866 | 3738 | 2543 | 665 | 789 |
| 1969 | 283767 | 90491 | 54478 | 31103 | 20546 | 2848 | 1970 | 832 |
| 1970 | 292893 | 215642 | 49165 | 35409 | 18699 | 11735 | 1808 | 1644 |
| 1971 | 354179 | 190991 | 118938 | 24641 | 19442 | 11948 | 7845 | 2527 |
| 1972 | 224067 | 209003 | 102729 | 67251 | 14508 | 11379 | 7314 | 6521 |
| 1973 | 201256 | 111280 | 105107 | 63144 | 35705 | 8679 | 6341 | 8626 |
| 1974 | 199522 | 90691 | 48306 | 62588 | 41888 | 20497 | 5397 | 8547 |
| 1975 | 234962 | 76745 | 35444 | 24189 | 36071 | 25047 | 11950 | 8493 |
| 1976 | 405133 | 102883 | 29664 | 17380 | 12852 | 19037 | 13210 | 11555 |
| 1977 | 149577 | 148834 | 35651 | 12423 | 8657 | 7176 | 10698 | 13931 |
| 1978 | 120502 | 72360 | 58123 | 14222 | 5106 | 4000 | 3381 | 13098 |
| 1979 | 87444 | 53914 | 34796 | 29302 | 7798 | 2810 | 2207 | 9524 |
| 1980 | 85530 | 47002 | 25721 | 18732 | 16090 | 4034 | 1666 | 7680 |
| 1981 | 162536 | 41815 | 24883 | 12288 | 9626 | 8277 | 2138 | 5895 |
| 1982 | 140877 | 108200 | 22975 | 15059 | 6273 | 4813 | 3759 | 4072 |
| 1983 | 148454 | 69524 | 54870 | 11340 | 8270 | 3131 | 2519 | 3813 |
| 1984 | 255733 | 76151 | 29972 | 23833 | 4723 | 3472 | 1328 | 2782 |
| 1985 | 356446 | 108526 | 29471 | 12768 | 9452 | 2220 | 1590 | 2299 |
| 1986 | 289571 | 142393 | 32268 | 11778 | 6363 | 4475 | 1192 | 2262 |
| 1987 | 149017 | 163930 | 36324 | 10186 | 5136 | 3286 | 2287 | 1801 |
| 1988 | 137951 | 71414 | 61486 | 11410 | 4548 | 2593 | 1741 | 1931 |
| 1989 | 102546 | 69331 | 27663 | 21785 | 4703 | 2090 | 1243 | 1649 |
| 1990 | 151294 | 48055 | 25651 | 11118 | 8374 | 2307 | 1026 | 1408 |
| 1991 | 174125 | 71728 | 17404 | 10236 | 5256 | 3795 | 1237 | 1393 |
| 1992 | 103944 | 88437 | 26152 | 6814 | 5170 | 2852 | 2059 | 1483 |
| 1993 | 176170 | 59297 | 33829 | 9173 | 2924 | 3131 | 1803 | 2259 |
| 1994 | 118110 | 97290 | 28687 | 13381 | 3461 | 1415 | 1468 | 2153 |
| 1995 | 215385 | 66737 | 41948 | 13018 | 6409 | 1626 | 924 | 1926 |
| 1996 | 119594 | 148907 | 29844 | 19650 | 7022 | 2502 | 711 | 1344 |
| 1997 | 148952 | 79763 | 90043 | 13262 | 9285 | 3434 | 1121 | 977 |
| 1998 | 87723 | 120609 | 45472 | 48923 | 7258 | 4639 | 1876 | 1051 |
| 1999 | 111435 | 55318 | 73745 | 22656 | 26645 | 4242 | 2389 | 1621 |
| 2000 | 97569 | 93077 | 29145 | 36652 | 11023 | 12662 | 2018 | 1741 |
| 2001 | 205817 | 67645 | 64004 | 14178 | 17604 | 6338 | 6919 | 1624 |
| 2002 | 163051 | 144410 | 35250 | 34548 | 8264 | 9667 | 3869 | 5009 |
| 2003 | 166920 | 123163 | 84968 | 16953 | 17419 | 5312 | 5279 | 4971 |
| 2004 | 117267 | 109637 | 75898 | 49011 | 8134 | 8300 | 3297 | 4664 |
| 2005 | 143615 | 75797 | 67544 | 50201 | 27953 | 5015 | 4533 | 4336 |
| 2006 | 102098 | 123653 | 42273 | 37337 | 26978 | 14160 | 3083 | 4859 |
| 2007 | 155694 | 55604 | 77448 | 24534 | 19864 | 14928 | 7164 | 4360 |
| 2008 | 73631 | 98547 | 31326 | 48339 | 15436 | 11156 | 10003 | 8123 |
| 2009 | 58640 | 51468 | 43732 | 14673 | 24714 | 9554 | 5849 | 10255 |
| 2010 | 90077 | 38225 | 28299 | 20281 | 7153 | 13075 | 5547 | 9542 |
| 2011 | 83142 | 79213 | 22793 | 14433 | 10139 | 3685 | 6511 | 9913 |
| 2012 | 141320 | 48196 | 48019 | 12244 | 7547 | 5027 | 1995 | 9212 |
| 2013 | 99244 | 107991 | 23823 | 27572 | 7055 | 4095 | 2676 | 6408 |
| 2014 | 60074 | 74017 | 60189 | 13739 | 15563 | 4201 | 2124 | 5287 |
| 2015 | 107275 | 44478 | 51242 | 34913 | 8110 | 8579 | 2696 | 4373 |

Table 11.5.1. Saithe in Subareas 4 and 6 and Division 3.a. Estimated recruitment, total stock biomass (TSB), spawning stock biomass (SSB), and average fishing mortality for ages 4 to 7 ( $\mathrm{F}_{4-7}$ ), 1967-2015. Low and High refer to the lower and upper $95 \%$ confidence interval estimates.

| Year | Recruits | Low | High | TSB | Low | High | SSB | Low | High | F47 | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 141049 | 100250 | 198451 | 413262 | 338348 | 504762 | 152998 | 120865 | 193674 | 0.351 | 0.273 | 0.451 |
| 1968 | 160710 | 116066 | 222527 | 579708 | 477847 | 703282 | 211139 | 169304 | 263312 | 0.292 | 0.229 | 0.374 |
| 1969 | 283767 | 204666 | 393440 | 711592 | 589975 | 858279 | 277606 | 225538 | 341693 | 0.320 | 0.257 | 0.399 |
| 1970 | 292893 | 212488 | 403722 | 909561 | 761892 | 1085851 | 346525 | 286158 | 419627 | 0.344 | 0.279 | 0.425 |
| 1971 | 354179 | 259600 | 483214 | 1054929 | 892982 | 1246247 | 461416 | 382154 | 557118 | 0.374 | 0.306 | 0.458 |
| 1972 | 224067 | 165369 | 303600 | 958361 | 819372 | 1120927 | 489786 | 408228 | 587638 | 0.406 | 0.334 | 0.493 |
| 1973 | 201256 | 148578 | 272612 | 893771 | 770064 | 1037351 | 521616 | 434824 | 625732 | 0.431 | 0.357 | 0.520 |
| 1974 | 199522 | 147024 | 270768 | 925363 | 802180 | 1067462 | 576364 | 483023 | 687744 | 0.498 | 0.417 | 0.595 |
| 1975 | 234962 | 174191 | 316934 | 856816 | 742955 | 988126 | 516418 | 431858 | 617535 | 0.537 | 0.451 | 0.640 |
| 1976 | 405133 | 294824 | 556715 | 812452 | 695827 | 948624 | 398488 | 331298 | 479304 | 0.601 | 0.504 | 0.717 |
| 1977 | 149577 | 110052 | 203296 | 612857 | 526548 | 713314 | 324904 | 269650 | 391480 | 0.592 | 0.491 | 0.714 |
| 1978 | 120502 | 88940 | 163265 | 520026 | 446249 | 606000 | 297101 | 245483 | 359575 | 0.482 | 0.401 | 0.580 |
| 1979 | 87444 | 64322 | 118876 | 483643 | 417009 | 560924 | 278838 | 232998 | 333698 | 0.452 | 0.375 | 0.544 |
| 1980 | 85530 | 62915 | 116273 | 439786 | 381075 | 507543 | 261217 | 219895 | 310303 | 0.474 | 0.397 | 0.567 |
| 1981 | 162536 | 118618 | 222716 | 492231 | 424278 | 571068 | 249741 | 211263 | 295227 | 0.473 | 0.395 | 0.566 |
| 1982 | 140877 | 104073 | 190695 | 530535 | 456292 | 616858 | 220175 | 188876 | 256661 | 0.542 | 0.459 | 0.640 |
| 1983 | 148454 | 109546 | 201183 | 508661 | 439507 | 588696 | 219844 | 188123 | 256914 | 0.652 | 0.553 | 0.769 |
| 1984 | 255733 | 188128 | 347633 | 515910 | 442169 | 601949 | 188332 | 161822 | 219185 | 0.677 | 0.578 | 0.794 |
| 1985 | 356446 | 259219 | 490140 | 528538 | 445297 | 627341 | 165787 | 143176 | 191968 | 0.702 | 0.600 | 0.822 |
| 1986 | 289571 | 213267 | 393176 | 492068 | 418570 | 578471 | 156686 | 135587 | 181069 | 0.736 | 0.623 | 0.869 |
| 1987 | 149017 | 109846 | 202158 | 403968 | 348185 | 468688 | 165372 | 143091 | 191122 | 0.706 | 0.602 | 0.828 |
| 1988 | 137951 | 102124 | 186347 | 348791 | 302157 | 402623 | 154711 | 132423 | 180750 | 0.713 | 0.608 | 0.836 |
| 1989 | 102546 | 75795 | 138739 | 292599 | 253392 | 337871 | 126303 | 108491 | 147039 | 0.691 | 0.588 | 0.812 |
| 1990 | 151294 | 111597 | 205111 | 301775 | 258308 | 352555 | 114398 | 98053 | 133468 | 0.657 | 0.558 | 0.772 |
| 1991 | 174125 | 128779 | 235437 | 320712 | 272751 | 377107 | 107444 | 92600 | 124667 | 0.627 | 0.533 | 0.738 |
| 1992 | 103944 | 77394 | 139602 | 310066 | 265901 | 361567 | 113008 | 97945 | 130387 | 0.608 | 0.514 | 0.718 |
| 1993 | 176170 | 130913 | 237072 | 356091 | 303049 | 418417 | 119604 | 103024 | 138852 | 0.640 | 0.540 | 0.757 |
| 1994 | 118110 | 87926 | 158656 | 339647 | 290514 | 397090 | 124892 | 107626 | 144928 | 0.569 | 0.480 | 0.674 |
| 1995 | 215385 | 158167 | 293303 | 452910 | 380876 | 538567 | 144156 | 123579 | 168159 | 0.574 | 0.482 | 0.684 |
| 1996 | 119594 | 88101 | 162346 | 433610 | 367185 | 512051 | 156161 | 134128 | 181813 | 0.507 | 0.424 | 0.606 |
| 1997 | 148952 | 108749 | 204016 | 448334 | 381481 | 526903 | 194461 | 164217 | 230273 | 0.440 | 0.365 | 0.530 |
| 1998 | 87723 | 64226 | 119818 | 394657 | 338703 | 459856 | 191216 | 162071 | 225601 | 0.448 | 0.373 | 0.537 |
| 1999 | 111435 | 81226 | 152879 | 380582 | 327806 | 441854 | 200392 | 169437 | 237003 | 0.480 | 0.398 | 0.580 |
| 2000 | 97569 | 71444 | 133248 | 396396 | 340283 | 461762 | 190854 | 161980 | 224874 | 0.417 | 0.344 | 0.507 |
| 2001 | 205817 | 150447 | 281565 | 449464 | 381999 | 528844 | 197732 | 167145 | 233915 | 0.385 | 0.316 | 0.471 |
| 2002 | 163051 | 119401 | 222659 | 495044 | 421179 | 581864 | 222354 | 188292 | 262579 | 0.399 | 0.330 | 0.483 |
| 2003 | 166920 | 122148 | 228102 | 448660 | 383297 | 525170 | 214089 | 181229 | 252908 | 0.413 | 0.341 | 0.500 |
| 2004 | 117267 | 86123 | 159673 | 505127 | 433222 | 588965 | 270812 | 228497 | 320963 | 0.362 | 0.296 | 0.442 |
| 2005 | 143615 | 104908 | 196603 | 481555 | 414121 | 559969 | 261942 | 221625 | 309592 | 0.373 | 0.307 | 0.453 |
| 2006 | 102098 | 73248 | 142312 | 501405 | 432072 | 581862 | 273621 | 231611 | 323252 | 0.388 | 0.320 | 0.470 |
| 2007 | 155694 | 111085 | 218216 | 463624 | 397626 | 540575 | 250461 | 211198 | 297023 | 0.369 | 0.303 | 0.448 |
| 2008 | 73631 | 53979 | 100438 | 435243 | 373892 | 506662 | 253402 | 213685 | 300502 | 0.436 | 0.359 | 0.529 |
| 2009 | 58640 | 42995 | 79978 | 394907 | 338948 | 460104 | 247285 | 207172 | 295164 | 0.441 | 0.363 | 0.535 |
| 2010 | 90077 | 65627 | 123636 | 400029 | 341318 | 468838 | 232304 | 192935 | 279706 | 0.424 | 0.348 | 0.517 |
| 2011 | 83142 | 59367 | 116439 | 362141 | 305198 | 429708 | 186024 | 153989 | 224722 | 0.426 | 0.345 | 0.525 |
| 2012 | 141320 | 99521 | 200674 | 375622 | 308204 | 457788 | 169156 | 138898 | 206007 | 0.385 | 0.305 | 0.486 |
| 2013 | 99244 | 67079 | 146832 | 386907 | 308665 | 484982 | 179454 | 144785 | 222425 | 0.342 | 0.260 | 0.451 |
| 2014 | 60074 | 37288 | 96785 | 390110 | 297856 | 510939 | 210261 | 163439 | 270496 | 0.309 | 0.220 | 0.434 |
| 2015 | 107275 | 56905 | 202231 | 417385 | 290382 | 599934 | 228761 | 165879 | 315480 | 0.301 | 0.197 | 0.460 |

Table 11.7.1. Saithe in Subareas 4 and 6 and Division 3.a. The basis for the catch options.

| VARIABLE | VALUE | Source | Notes |
| :--- | :---: | :--- | :--- |
| F ages 4-7 (2016) | 68601 t | ICES <br> (2016a) | TAC constraint (F=0.24) |
| SSB (2016) | 239561 t | ICES <br> (2016a) | SSB in the intermediate year |
| SSB (2017) | 277948 t | ICES <br> (2016a) | SSB at the beginning of the TAC year |
| Rage3 (2016) | 102 billion | ICES <br> (2016a) | Median recruitment re-sampled from 2003- <br> 2015 |
| Rage3 (2017) | 102 billion | ICES <br> $(2016 a)$ | Median recruitment re-sampled from 2003- <br> 2015 |
| Total catch (2016) | 72442 t | ICES <br> (2016a) | Assuming 2015 landings fraction by age |
| Commercial <br> landings (2016) | 68601 t | ICES <br> (2016a) | TAC 2015 |
| Discards (2016) | 3841 t | ICES <br> (2016a) | Assuming 2015 discard fraction by age |

* TAC was based on landings without adjustment.

Table 11.7.2. Saithe in Subareas 4 and 6 and Division 3.a. All weights in tonnes.

| RATIONALE | TOTAL CATCH (2017) | Wanted CATCH* (2017) | UNWANTED CATCH* (2017) | WANTED <br> CATCH $\begin{aligned} & 3 . A \& 4 \\ & 2017 \text { ** } \end{aligned}$ | Wanted $\begin{gathered} \text { CATCH } \\ 6 \\ 2017 \text { ** } \end{gathered}$ | BASIS | $\begin{aligned} & \text { FTOTAL } \\ & \text { (2017) } \end{aligned}$ | FWANTED (2017) | Funwanted (2017) | $\begin{gathered} \text { SSB } \\ (2018) \end{gathered}$ | $\begin{gathered} \text { \%SSB } \\ \text { CHANGE } \\ * * \end{gathered}$ | \% TAC <br> Change <br> WANTED <br> CATCH^ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MSY approach | 116605 | 110917 | 5688 | 100491 | 10426 | FMSY | 0.36 | 0.34 | 0.02 | 278294 | 0 | 62 |
| EU-Norway management strategy | 82747 | 78850 | 3897 | 71438 | 7412 | Paragraph 5 of management strategy | 0.24 | 0.23 | 0.01 | 310717 | 12 | 15 |
| Zero catch | 0 | 0 | 0 | 0 | 0 | $\mathrm{F}=0$ | 0 | 0 | 0 | 390077 | 40 | -100 |
| Other options | 81833 | 77982 | 3851 | 70652 | 7330 | F2016 | 0.24 | 0.23 | 0.01 | 311578 | 12 | 14 |
|  | 72171 | 68601 | 3570 | 62153 | 6448 | TAC2016 | 0.21 | 0.2 | 0.01 | 321351 | 16 | 0 |
|  | 128157 | 121919 | 6237 | 110459 | 11460 | Fpa | 0.40 | 0.38 | 0.02 | 267579 | -4 | 78 |
|  | 167792 | 159522 | 8270 | 144527 | 14995 | Flim | 0.56 | 0.54 | 0.02 | 229799 | -17 | 133 |
|  | 73497 | 70040 | 3457 | 63456 | 6584 | Flower (w/AR) | 0.21 | 0.2 | 0.01 | 319442 | 15 | 2 |
|  | 128157 | 121919 | 6237 | 110459 | 11460 | Fupper (w/AR) | 0.49 | 0.47 | 0.02 | 246040 | -11 | 109 |
|  | 304748 | 287277 | 17470 | 260273 | 27004 | SSB2018 = Blim | 1.40 | 1.33 | 0.07 | 107000 | -62 | 319 |
|  | 255135 | 241105 | 14030 | 218441 | 22664 | SSB2018 $=$ Bpa | 1.02 | 0.98 | 0.04 | 150000 | -46 | 251 |
|  | 255135 | 241105 | 14030 | 218441 | 22664 | $\begin{aligned} & \text { SSB2018= } \\ & \text { Btrigger*** } \end{aligned}$ | 1.02 | 0.98 | 0.04 | 150000 | -46 | 251 |
|  |  |  |  |  |  | F in 0.01 increments (0.22) |  |  |  |  |  |  |
|  | 75678 | 72117 | 3560 | 65338 | 6779 |  | 0.22 | 0.21 | 0.01 | 317383 | 14 | 5 |
|  | 78769 | 75063 | 3706 | 68007 | 7056 | $\mathrm{F}=0.23$ | 0.23 | 0.22 | 0.01 | 314466 | 13 | 9 |
|  | 81833 | 77982 | 3851 | 70652 | 7330 | $\mathrm{F}=0.24$ | 0.24 | 0.23 | 0.01 | 311578 | 12 | 14 |
|  | 84869 | 80865 | 4004 | 73264 | 7601 | $\mathrm{F}=0.25$ | 0.25 | 0.24 | 0.01 | 308717 | 11 | 18 |
|  | 87877 | 83722 | 4155 | 75852 | 7870 | $\mathrm{F}=0.26$ | 0.26 | 0.25 | 0.01 | 305884 | 10 | 22 |
|  | 90859 | 86552 | 4307 | 78416 | 8136 | $\mathrm{F}=0.27$ | 0.27 | 0.26 | 0.01 | 303073 | 9 | 26 |
|  | 93830 | 89356 | 4474 | 80957 | 8399 | $\mathrm{F}=0.28$ | 0.28 | 0.27 | 0.01 | 300222 | 8 | 30 |
|  | 96777 | 92134 | 4644 | 83473 | 8661 | $\mathrm{F}=0.29$ | 0.29 | 0.28 | 0.01 | 297377 | 7 | 34 |
|  | 99699 | 94891 | 4809 | 85971 | 8920 | $\mathrm{F}=0.30$ | 0.3 | 0.29 | 0.01 | 294542 | 6 | 38 |
|  | 102597 | 97623 | 4973 | 88446 | 9177 | $\mathrm{F}=0.31$ | 0.31 | 0.3 | 0.01 | 291762 | 5 | 42 |
|  | 105450 | 100331 | 5119 | 90900 | 9431 | $\mathrm{F}=0.32$ | 0.32 | 0.31 | 0.01 | 289031 | 4 | 46 |
|  | 108276 | 103014 | 5262 | 93331 | 9683 | $\mathrm{F}=0.33$ | 0.33 | 0.31 | 0.02 | 286287 | 3 | 50 |
|  | 111078 | 105672 | 5405 | 95739 | 9933 | $\mathrm{F}=0.34$ | 0.34 | 0.32 | 0.02 | 283557 | 2 | 54 |
|  | 113854 | 108307 | 5547 | 98126 | 10181 | $\mathrm{F}=0.35$ | 0.35 | 0.33 | 0.02 | 280912 | 1 | 58 |
|  | 116605 | 110917 | 5688 | 100491 | 10426 | $\mathrm{F}=0.36$ | 0.36 | 0.34 | 0.02 | 278294 | 0 | 62 |
|  | 119331 | 113503 | 5828 | 102834 | 10669 | $\mathrm{F}=0.37$ | 0.37 | 0.35 | 0.02 | 275701 | -1 | 65 |
|  | 122033 | 116079 | 5954 | 105168 | 10911 | $\mathrm{F}=0.38$ | 0.38 | 0.36 | 0.02 | 273161 | -2 | 69 |
|  | 124711 | 118633 | 6078 | 107481 | 11152 | $\mathrm{F}=0.39$ | 0.39 | 0.37 | 0.02 | 270719 | -3 | 73 |
|  | 127365 | 121164 | 6201 | 109775 | 11389 | $\mathrm{F}=0.40$ | 0.4 | 0.38 | 0.02 | 268300 | -3 | 77 |
|  | 129995 | 123673 | 6322 | 112048 | 11625 | $\mathrm{F}=0.41$ | 0.41 | 0.39 | 0.02 | 265886 | -4 | 80 |
|  | 132602 | 126160 | 6442 | 114301 | 11859 | $\mathrm{F}=0.42$ | 0.42 | 0.4 | 0.02 | 263414 | -5 | 84 |
|  | 135185 | 128625 | 6560 | 116534 | 12091 | $\mathrm{F}=0.43$ | 0.43 | 0.41 | 0.02 | 260899 | -6 | 87 |
|  | 137745 | 131061 | 6684 | 118741 | 12320 | $\mathrm{F}=0.44$ | 0.44 | 0.42 | 0.02 | 258408 | -7 | 91 |
|  | 140290 | 133465 | 6824 | 120919 | 12546 | $\mathrm{F}=0.45$ | 0.45 | 0.43 | 0.02 | 255942 | -8 | 95 |
|  | 142812 | 135847 | 6965 | 123077 | 12770 | $\mathrm{F}=0.46$ | 0.46 | 0.44 | 0.02 | 253500 | -9 | 98 |
|  | 145313 | 138208 | 7105 | 125216 | 12992 | $\mathrm{F}=0.47$ | 0.47 | 0.45 | 0.02 | 251113 | -10 | 101 |
|  | 147791 | 140546 | 7245 | 127335 | 13211 | $\mathrm{F}=0.48$ | 0.48 | 0.46 | 0.02 | 248794 | -10 | 105 |

[^5]** Wanted catch split according to the average in 1993-1998, i.e. $90.6 \%$ in Subarea 4 and Subdivision 3.a.20 and 9.4\% in Subarea
6.
*** $\mathbf{B}_{\text {trigger }}=\mathbf{B}_{\mathrm{pa}}$
^ Wanted catch 2017 relative to the 2016 wanted catch (without adjustment) TAC.

Table 11.7.3. Saithe in Subareas 4 and 6 and Division 3.a. Contribution of the year classes to the landings in 2017.

|  | Year class | Contribution to Landings (\%) |
| :--- | :--- | :--- |
| 2014 | 7 |  |
| 2013 | 20 |  |
| 2012 | 23 |  |
| 2011 | 10 |  |
| 2010 | 14 |  |
| 2009 | 12 |  |
| 2008 | 4 |  |

Table 11.9.1. Saithe in Subareas 4 and 6 and Division 3.a. Reference points estimated using the FBI + DATRAS Q3 model, where catch weights = stock weights for all ages.

|  | Reference point |
| :--- | :--- |
| Blim | 107000 |
| Bpa (1.4) | 150000 |
| Bpa (sigma) | 139000 |
| Btrigger | 150000 |
| Flim | 0.564 |
| Fpa (1.4) | 0.403 |
| Fpa (sigma) | 0.398 |
| FMSY without Btrigger | 0.36 |
| FMSY lower without Btrigger | 0.21 |
| FMSY upper without Btrigger | 0.498 |
| New FP.05 (5\% risk to Blim without Btrigger) | 0.419 |
| FMSY upper precautionary without Btrigger | 0.419 |
| FP.05 (5\% risk to Blim with Btrigger) | 0.492 |
| FMSY with Btrigger | 0.395 |
| FMSY lower with Btrigger | 0.213 |
| FMSY upper with Btrigger | 0.647 |
| FMSY upper precautionary with Btrigger | 0.492 |
| MSY (without HCR) | 89305 |
| Median SSB at FMSY (without HCR) | 206513 |
| Median SSB lower precautionary (median at FMSY upper precautionary; |  |
| without HCR) | 179497 |
| Median SSB upper (median at FMSY lower; without HCR) | 368806 |

Sigma $(F)=0.2114521$, sigma $(S S B)=0.1607088$.

Table 11.10.1. Saithe in Subareas 4 and 6 and Division 3.a. Model and data selection settings

| DATA AND PARAMETERS | SETTING | CommENTS |
| :--- | :--- | :--- |
| Recruitment model | $\begin{array}{l}\text { Segmented } \\ \text { regression, where the } \\ \text { inflection point was } \\ \text { forced to be Bloss } \\ \text { from the entire time } \\ \text { series }\end{array}$ | $\begin{array}{l}\text { Recruitment vs. SSB for the entire times series } \\ \text { showed a distinct plateau across a wide range of } \\ \text { SSB. For stocks showing this characteristic, Bloss } \\ \text { is recommended to be the inflection point in the } \\ \text { segmented regression. }\end{array}$ |
| $\begin{array}{ll}\text { SSB-recruitment } \\ \text { data }\end{array}$ | $\begin{array}{l}\text { (a) Truncated time } \\ \text { series, based on } \\ \text { changepoint analysis } \\ \text { (year classes 2000 to } \\ 2011)\end{array}$ | $\begin{array}{l}\text { Changepoint analysis of R per SSB showed } \\ \text { distinct periods in recruitment: higher R per SSB } \\ \text { in 1969-2002 and lower in 2003 to present (see } \\ \text { also section sensitivity/discussion). }\end{array}$ | \(\left.\begin{array}{lll}(b) Full data series <br>

(year classes 1967 to <br>
2011)\end{array} \quad $$
\begin{array}{l}\text { R per SSB shows signs of cyclic changes in } \\
\text { productivity over time. Whether the current low } \\
\text { productivity of the stock can be explained by }\end{array}
$$\right\}\)


Figure 11.3.1. Saithe in Subareas 4 and 6 and Division 3.a. Reported landings for each sampled and unsampled fleet in the full stock area, along with cumulative landings for fleets in descending order of yield.


Figure 11.3.2. Saithe in Subareas 4 and 6 and Division 3.a. Overview of percent of catch sampled and unsampled by country, fleet, and quarter for saithe catches in Subdvision 3.a.


Figure 11.3.3. Saithe in Subareas 4 and 6 and Division 3.a. Overview of percent of catch sampled and unsampled catches by country, fleet, and quarter for saithe catches in Subarea 4. Scotland reported by year, not quarter.



Figure 11.3.5. Saithe in Subareas 4 and 6 and Division 3.a. Summary of landings for fleets with and without discard estimates.


Figure 11.3.6. Saithe in Subareas 4 and 6 and Division 3.a. Proportion of total catch discarded by age and year.


Figure 11.3.7. Saithe in Subareas 4 and 6 and Division 3.a. Yield by catch component.


Figure 11.3.8. Saithe in Subareas 4 and 6 and Division 3.a. (left) Landings-at-age for saithe ages 310+, 1990-2015; smallest bubble corresponds to 210 thousand individuals and largest to 46 million individuals. (right) Discard weights at age for saithe ages 3-10+, 2000-2015 (min: 0, max: 15 million individuals).


Figure 11.3.9. Saithe in Subareas 4 and 6 and Division 3.a. (left) Catch weight-at-age (kg) for saithe ages $3-10+$, 1967-2015. Catch weight-at-age are also stock weight-at-age in the assessment. (right) Discard weights-at-age (kg) for saithe ages 3-10+, 1967-2015.


Figure 11.4.1. Saithe in Subareas 4 and 6 and Division 3.a. Log catch curves by cohort from the research survey index, IBTS Q3, for total catches for ages 3 to 8 .


Figure 11.4.2. Saithe in Subareas 4 and 6 and Division 3.a. Internal consistencies for IBTS Q3, ages 3 to 8.


Figure 11.4.3. Saithe in Subareas 4 and 6 and Division 3.a. Standardised IBTS Q3 research tuning series index.


Figure 11.4.4. Saithe in Subareas 4 and 6 and Division 3.a. Log catch curves by cohort for landings for ages 3 to 9 .


Figure 11.4.5. Saithe in Subareas 4 and 6 and Division 3.a. Standardized combined cpue index (year effects) and fit of model after tuning to the biomass. Left: Fit for the exploratory assessment model: combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for ages 7+. Right: Fit for the final model: combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for all ages.


Figure 11.4.6. Saithe in Subareas 4 and 6 and Division 3.a. Fishing mortality at age. Left: Assessment model is combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for ages 7+. Right: Assessment model is combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for all ages.


Figure 11.4.7. Saithe in Subareas 4 and 6 and Division 3.a. Residual patterns for the exploratory assessment model: combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for ages $7+$. (left) Before correlation taken into account between ages; (right) after accounting for the correlation between ages within years (residuals are one-step ahead). Open circles (blue) indicate positive residuals and filled red circles indicate negative residuals.


Figure 11.4.8. Saithe in Subareas 4 and 6 and Division 3.a. Eight year retrospective pattern in SSB, $\mathrm{F}_{4}-7$, and recruitment for the exploratory model: combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for ages $7+$.


Figure 11.4.9. Saithe in Subareas 4 and 6 and Division 3.a. Stock summary of trends in SSB, F4-7, and recruitment for the exploratory model: combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for ages $7+$. Blue line and grey dashed confidence interval pertains to the final assessment model: combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for all ages.


Figure 11.4.10. Saithe in Subareas 4 and 6 and Division 3.a. Residual patterns for the final SAM model: combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for all ages. Left: Before correlation taken into account between ages, within years in the Q3 index. Right: After accounting for the correlation between ages within years in the Q3 index. Open circles (blue) indicate positive residuals and filled red circles indicate negative residuals.


Figure 11.4.11. Saithe in Subareas 4 and 6 and Division 3.a. Eight year retrospective pattern in SSB, $\mathrm{F}_{4-7}$, and recruitment for the final assessment model: combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for all ages.


Figure 11.5.1. Saithe in Subareas 4 and 6 and Division 3.a. Stock summary of trends in SSB, F4-7, and recruitment for the final assessment model: combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat), stock weights=catch weights for all ages.


Figure 11.9.1. Saithe in Subareas 4 and 6 and Division 3.a. Left: Stock recruitment relationship based on segmented regression over the truncated time period (2003-2015), where the inflection point was forced to be Bloss from the complete time series 1967-2014. Right: Stock recruitment relationship based on segmented regression over the entire time series.


Figure 11.9.2. Saithe in Subareas 4 and 6 and Division 3.a. Recruitment per SSB over time and periods identified in the change-point analysis as having different levels of recruitment.


Figure 11.10.1. Saithe in Subareas 4 and 6 and Division 3.a. Eqsim results: no trim, no $B_{\text {trigger, }}$, segmented regression with $B_{l o s s}$ forced to be the inflection point. Panels a-c: historic values (dots), median (solid black), and $90 \%$ intervals (dotted black) for recruitment, SSB, and landings for exploitation at fixed values of $F$. Panel calso shows mean landings (red solid line). Panel d shows the probability of $S S B<B_{\lim }$ (red), $S S B<B_{p a}$ (green), and the cumulative distribution of $F_{m S y}$ based on yield as landings (brown) and catch (cyan).


Figure 11.10.2. Saithe in Subareas 4 and 6 and Division 3.a. Eqsim results: no trim, with $B_{\text {trigger, }}$ segmented regression with $B_{\text {loss }}$ forced to be the inflection point). Panels a-c: historic values (dots), median (solid black), and $90 \%$ intervals (dotted black) for recruitment, SSB, and landings for exploitation at fixed values of F. Panel c also shows mean landings (red solid line). Panel d shows the probability of $S S B<B_{\lim }$ (red), $S S B<B_{\text {pa }}$ (green), and the cumulative distribution of $F_{m s y}$ based on yield as landings (brown) and catch (cyan).


Figure 11.10.3. Saithe in Subareas 4 and 6 and Division 3.a. Eqsim results with a fixed $F$ exploitation from F = 0 to 1.2 (no $B_{\text {trigger }}$ ). Left: median landing yield curve: blue lines: $\mathrm{F}_{\mathrm{msy}}$ estimate (solid) and range at $95 \%$ of maximum yield, with the upper bound restricted to $\mathrm{F}_{\mathrm{F} .05}$ (dotted); green lines: $\mathrm{F}(5 \%$ ) estimate (solid). Right: median SSB: blue lines show the location of Fmsy (solid) with the (dotted) lower $\mathbf{9 5 \%}$ Fmsy and the upper precautionary bound (restricted to $\mathrm{Fp}_{\mathrm{p} .05}$ ).


Figure 11.10.4. Saithe in Subareas 4 and 6 and Division 3.a. Eqsim results with a fixed $F$ exploitation from F = $\mathbf{0}$ to 1.2 and applying the ICES MSY harvest control rule with a $B_{\text {trigger }}$ at 150000 tonnes. Left: median landing yield curve: blue lines: Fmš estimate (solid) and range at $95 \%$ of maximum yield, with the upper bound restricted to Fpo.05 (dotted); green lines: F(5\%) estimate (solid). Right: median SSB: blue lines show the location of $\mathrm{F}_{\text {ms }}$ (solid) with the (dotted) lower $95 \%$ Fmš and the upper precautionary bound (restricted to Fp .05 ).

## 12 Whiting (Merlangius merlangus) in Subarea 4 (North Sea), Division 7.d (Eastern English Channel) and 3.a (Skagerrak and Kattegat)

### 12.1 Whiting in Subarea 4 and Divisions 7.d

This Section contains the assessment relating to whiting in the North Sea (ICES Subarea 4) and eastern Channel (ICES Division 7.d). The current assessment is formally classified as an update assessment. The most recent benchmark for this stock was conducted in January 2013 (WKROUND 2013). An Interbenchmark was conducted in March 2016 (ICES 2016) to test new natural mortalities from the 2014/2015 key run from of the SMS multispecies model (WGSAM 2014).
Available information on whiting in Division 3.a (Skagerrak and Kattegat) is presented in Section 12.2.

### 12.1.1 General

### 12.1.1.1 Stock definition

No new information was presented at the WG. A summary of available information on stock definition can be found in the Stock Annex prepared by ICES-WKROUND (2013).

### 12.1.1.2 Ecosystem aspects

No new information was presented at the WG. A summary of available information on ecosystem aspects is presented in the Stock Annex prepared by ICES-WKROUND (2013).

### 12.1.2 Fisheries

Information on the fishery (and its historical development) is contained in the Stock Annex prepared by ICES-WKROUND (2013).

### 12.1.3 ICES advice

ICES advice for 2014
In November 2013, ICES concluded as follows:
ICES advises on the basis of precautionary considerations that total catches should be no more than 31553 tonnes. If rates of discards and industrial bycatch do not change from the average of the last three years (2010-2012), this implies human consumption landings of no more than 21199 tonnes ( 16092 tonnes in the North Sea and 5106 tonnes in Division 7.d). Management for Division 7.d should be separated from the rest of Subarea 7.

## ICES advice for 2015

In November 2014, ICES concluded as follows:
ICES advised on the basis of the EU-Norway management plan that total catches should be no more than 30579 tonnes. If rates of discards and industrial bycatch do not change from the average of the last three years (2011-2013), this implies human consumption of no more than 17190 tonnes ( 13678 tonnes in the North Sea and

3512 tonnes in Division 7.d). Management for Division 7.d should be separated from the rest of Subarea 7.

## ICES advice for 2016

In November 2015, ICES concluded as follows:
ICES advised on the basis of the EU-Norway management plan that total catches should be no more than 30510 tonnes. If rates of discards and industrial bycatch do not change from the average of the last three years (2012-2014), this implies human consumption of no more than 14853 tonnes (12373 tonnes in the North Sea and 2480 tonnes in Division 7.d). Management for Division 7.d should be separated from the rest of Subarea 7.

### 12.1.4 Management

Management of whiting is by TAC and technical measures. The TACs for this stock are split between two areas: (i) Subarea 4 and Division 2.a (EU waters), and (ii) Divisions 7b-k. Since 1996 the North Sea and eastern Channel whiting assessments have been combined into one.

The agreed TACs for whiting in Subarea 4 and Division 2.a (EU waters) were 16092 t in 2014 and 13678 t in 2015. The TAC in 2016 was set as a Roll-over TAC at 13 678t. There is no separate TAC for Division 7.d; landings from this Division are counted against the TAC for Divisions 7b-k combined ( 20668 t in 2014, 17742 t in 2015 and 22778 t in 2016). There are no means to control how much of the Division 7b-k TAC is taken from Division 7.d. By comparison, a specific TAC for Division 7.d was established for cod in 2009, and the same procedure for whiting may be appropriate.

In previous years, the human consumption landings in Subarea 4 and Division 7.d were calculated as $70 \%$ and $30 \%$ of the combined area totals. Since 2006, the landings data have been collated separately for each area. In 2015, $77 \%$ of the total catches originated in Subarea 4.

The minimum landing size for whiting in Subarea 4 and Division $7 . \mathrm{d}$ is 27 cm . The minimum mesh size for whiting in Division 7.d is 80 mm .

Whiting are a by-catch in some Nephrops fisheries that use a smaller mesh size, although landings are restricted through by-catch regulations. They are also caught in flatfish fisheries that use a smaller mesh size. Industrial fishing with small-meshed gear is permitted, subject to by-catch limits of protected species. Regulations also apply to the area of the Norway pout box, preventing industrial fishing with small meshes in an area where the by-catch limits are likely to be exceeded.

## Conservation credit scheme

During 2008, 15 real-time closures (RTCs) were implemented under the Scottish Conservation Credits Scheme (CCS). In 2009, 144 RTCs were implemented, and the CCS was adopted by 439 Scottish and around 30 English and Welsh vessels. In 2010 there were 165 closures, and from July 2010 the area of each closure increased (from 50 square nautical miles to 225 square nautical miles). In more recent years, the following numbers of closures were implemented: 185 (2011), 173 (2012), 166 (2013), 94 (2014) and 97 (2015). In 2016, 34 closures had been implanted by 4th May. The CCS has two central themes aimed at reducing the capture of cod through (i) avoiding areas with elevated abundances of cod through the use of Real Time Closures (RTCs) and (ii) the use of more species selective gears. Within the scheme, efforts are also being made to reduce
discards generally. Although the scheme is intended to reduce mortality on cod, it will undoubtedly have an effect on the mortality of associated species such as whiting.
Studies tracking Scottish vessels during 2009-2010 concluded that vessels did indeed move from areas of higher to lower cod concentration following real-time closures during the first and third quarters, although there was no significant effect during the second and fourth quarters; see Needle and Catarino (2011). In a subsequent analysis, Needle (2012) showed that the net effect of RTCs appeared to be to attract vessels, although the movement towards RTCs may have been coincidental. However, the effect of these changes in behavior on the whiting stock is still under investigation.
In 2015, 24 Scottish demersal whitefish vessels (although 6 left during the year) participated in a trial Fully Documented Fishery (FDF) scheme, following similar schemes during 2010-2014. Trials of similar schemes have been conducted during various periods by Denmark, England, Germany, Sweden and the Netherlands. In the Scottish North Sea FDF scheme, vessels are exempt from some effort restrictions and are allocated additional cod quota: in return, they must carry monitoring cameras and land all cod caught. It is not clear what the impact would be on whiting fisheries of an enforceable discard ban for cod, and in data collation for the whiting assessment it was assumed that FDF vessels would have similar whiting discard patterns as other vessels, but this remains to be verified. The Scottish FDF scheme for 2016 began in January, and is being run along similar lines to previous years. Currently, 12 vessels are participating in the scheme in 2016. The uptake of the scheme has declined due to concerns about the monitoring of discards under the Landing Obligation.

### 12.1.5 Data available

### 12.1.5.1 Catch

Since 2012, international data on landings and discards have been collated through the InterCatch system.

The $70 \%$ of the landings had associated discard data imported to Intercatch. The discard data provided for landings in 2015 are illustrated in Figure 12.1. Discards were raised from discard ratios for all strata from Subarea 4 and Division 7.d combined. Industrial bycatch landings were excluded from the discard raising, as now discards occur in that fleet. Minor whiting bycatch landings of 12 t from a miscellaneous fleet (originating from a Dutch pelagic métier under landing obligation) imported as BMS landing (below minimum landing size) into InterCatch were treated as discards throughout.

Figure 12.2.1 shows fleet-specific landings in percent of the total landings in 2015 for whiting in Subarea 4 and Division 7.d, for fleets sampled for age compositions in landings and for fleets which were not sampled. The Figure also shows the cumulative landings when sampled and unsampled fleets are ordered by landings yield. Sampled fleets comprise around $69 \%$ of the overall landings, from 9 métiers. Sampled and unsampled métiers are listed in Figure 12.2.1. However, although the unsampled fleets provide considerable landings overall (30\%), most métiers provide less than $5 \%$ of the overall landings each. A métier summarized as miscellaneous landing industrial bycatch provides about $11 \%$ of the landings, occurred in the Danish fishery and was not sampled. It would therefore make little difference to the final data collation to segregate fleets on the basis of gear type or quarter before applying age compositions. Age compositions were applied to landings without any splitting of fleets on the basis of
area, quarter or gear type. For consistency, the same approach was taken for raising discard rates from sampled to unsampled discard fleets.
Of the total discards, $83 \%$ were imported into InterCatch. $59 \%$ of the imported discards were sampled for age distributions. The 11 métiers providing discard samples and unsampled métiers are listed in Figure 12.2.2.
Discard rates for unsampled whiting fleet components were obtained from samples provided by France and UK (England, Scotland) for Subarea 4 and in Division 7.d by France.

Official reported landings by country, WG estimates of total catch and catch component yields, as well as TACs covering the respective areas are given in Table 12.1 for the North Sea (Subarea 4) and in Table 12.2 for the Eastern Channel (Division 7.d).
WG estimates of numbers and weights at age for the defined catch components (total catch, landings, discards and industrial bycatch) are given in Tables 12.4 to 12.11. The estimated tonnages of the Subarea 4 catch components remained low but is higher than in most recent years, and whiting industrial by-catch remains low similar to last year ( $6 \%$ of the total catches). Discards have increased to $40 \%$ of the total catches, discards in both catch components increased. Figure 12.3 plots the trends in the commercial catch for each component in both Subarea 4 and Division 7.d combined. Recent years have seen these time series stabilize to a certain extent. There has been a slight increase in discards for age 1 in recent years. Compared to last year, the discard increased for all age groups except age 1, 4, 6 (Figure 12.4).

### 12.1.5.2 Age compositions

Age compositions in the landings and discards were based on samples provided by France, UK(England) and UK(Scotland). Limited sampling of the industrial bycatch component resulted in the 2006 data appearing as an outlier and the 2007 to 2010 data were deemed unreliable. This applies to both the age compositions and the estimates of mean weights at age. Thus the data for 2006 to 2010 were replaced with estimates derived from the years 1990 to 2005 (as described in the Stock Annex).

For the industrial bycatch in 2011 and 2012, age compositions were inferred in InterCatch from corresponding age samples taken from small-mesh fisheries of France and the UK.

Total international catch numbers at age (Subarea 4 and Division 7.d combined) are presented in Table 12.4. Numbers for human consumption landings, discards, and industrial bycatch are given in Tables 12.5 to 12.7.

### 12.1.5.3 Weight at age

Mean weights at age (Subarea 4 and Division $7 . d$ combined) in the catch are presented in Table 12.8. These are also used as stock weights at age. Mean weights at age (both areas combined) in human consumption landings are presented in Table 12.9, and for the discards and industrial by-catch in the North Sea in Tables 12.10 and 12.11 respectively. Weights-at-age are depicted graphically in Figure 12.5, which indicates an increasing trend (with annual fluctuations) in mean weight-at-age in the landings, discards and total catch for ages $>2$.

Unrepresentative sampling of industrial bycatch in 2006 to 2010 resulted in poor estimates of the mean weights at age and these have been replaced by the mean weight at age for the period 1995 to 2005 (zero weights are taken as missing values). For 2012 to

2015, the weights at ages of total catches were used for weights at ages of industrial bycatches.

### 12.1.5.4 Maturity and natural mortality

Values for maturity remain unchanged from those used in recent assessments and are given in Table 12.12. Their origin is discussed in the Stock Annex.

Estimates of natural mortality (M) are taken from the 2014/2015 update key run from of the SMS multispecies model (ICES-WGSAM 2014) (Table 12.13 and Figure 12.6). It was decided by WGNSSK to use the most recent estimates of natural mortality values from the 2014/2015 model key run, because recruitment estimates in the assessment changed significantly with the new estimates, while SSB and F were hardly impacted. The SMS keyrun is mainly based on stomach data sampled in the years 1981, 85-87 and 1991. In addition, data on the diet of marine mammals (seals, harbour porpoise) are available from sporadic samples during the last 30 years. In general, the new keyrun in is an update of the 2011 keyrun. But in comparison, the time series of grey gurnard and raja abundances were revised. In addition, the cod assessment changed between 2011 and 2014 leading to lower abundances of cod in SMS. The predation mortalities for age 1 and 2 are systematically lower than in the 2011 keyrun. The assessment results using the new natural mortalities questioned the current reference points and an Interbenchmark was done to test the new natural mortality estimates (ICES 2016). As a result, it was decided to use these new mortality values. The new natural mortality values are constant for ages 5+ (Figure 6). Reference points were updated accordingly (Table 12.19).

### 12.1.5.5 Research vessel data

Survey tuning indices are presented in Table 12.14. The indices used in the assessment are ages 1 to 5 from the IBTS-Q1 and IBTS-Q3 surveys, from 1990 to 2016 and 1991 to 2015, respectively. The report of the 2001 meeting of WGNSSK (ICES-WGNSSK 2002), and the ICES advice for 2002 (ICES-ACFM 2001) provide arguments for the exclusion of commercial CPUE tuning series from calibration of the catch-at-age analysis. Such arguments remain valid and only survey data have been considered for tuning purposes. All available tuning series are presented in the Stock Annex prepared at ICESWKROUND (2013).

In Figure 12.7 survey distribution maps, based on the IBTS-Q1 survey in the North Sea, for ages 1-3+ of the first quarter (Q1) 2012-2016 are presented. Figure 12.8, the third quarter is represented (Q3) for ages $0-3+$ for the years 2012-2015. The figures illustrate the CPUE is high along the UK east coast in quarter 1 and in the Northern North Sea and Scottish east coast and in the German Bight in quarter 3 for age 0-2. For age3+ the CPUE is highest along the UK East Coast. In 2015/2016 CPUE generally remained high as the last year.

### 12.1.6 Data analyses

The benchmark meeting for whiting in Subarea 4 and Division 7.d was held in Galway and Aberdeen in early 2013 (ICES-WKROUND 2013). Analyses focused on a number of key issues: these are listed below, along with relevant recommendations for future work (and steps taken by WGNSSK to address them):

## CCTV-based discard-rate estimation

Several participating countries have now installed CCTV cameras on a subset of vessels, and the issue is whether footage from these can be used to improve discard- rate estimation for assessments. The WKROUND meeting concluded that further work is needed to integrate CCTV with existing observer programs, and work is ongoing to improve length measurements accordingly (new camera and annotation systems, automated image analysis, and length-based assessments).

## Length of assessment time-series

Considerable effort was put into the evaluation of the pre-1990 catch and survey data which were previously used in the assessment, but which were removed in recent years due to discrepancies between catch and survey information. WKROUND found that pre-1990 catch data would need to be reduced by at least $75 \%$ for the FLXSA SSB estimates (catch-based) to resemble those from SURBAR (survey-based). It did not seem possible to resolve this discrepancy, and WKROUND concluded that 1990 should be retained as the starting point for the update assessment.

## Stock identity

The issue of how to define stock units for whiting that are biologically relevant remains a difficult one to address. WKROUND evaluated the available evidence, and produced area-specific SURBAR analyses to determine whether estimated time-series of biomass and mortality were correlated between different areas. Although the northern North Sea appeared to be linked with the areas immediately to the south and with no others, the analysis was not sufficiently conclusive. There is some evidence for north-south split in the North Sea, and some evidence for links between Divisions 4.a and 6.a (Holmes et al. 2014), but full stock determination is hindered by data availability. It would be very difficult to subdivide historical landings and discards time-series from all participating nations between any new areas. WKROUND 2013 concluded that the issue of stock identity needs to be considered as a matter of high priority, and as a parallel process with the existing data collation and assessment approach.

## Assessment models

WKROUND concluded that the update assessment model should continue to be FLXSA, with supporting exploratory runs using SURBAR (and, time permitting, SAM). A full investigation of the appropriate SAM run settings was not possible due to lack of time, although WKROUND recommended that this be done in the near future.

### 12.1.6.1 Exploratory survey-based analyses

Figure 12.9 presents time-series of survey $\log$ CPUE at age, and suggests that while broad trends are captured in a consistent way by the two surveys, finer-scale details of year-class strength may not be.

Catch-curve analyses for the surveys are shown in Figures 12.10. These show consistent tracking of year classes (since catch curves are mostly smooth) and consistent selection with some recent exceptions. The catchability of the IBTS Q1 seems to have changed since 2007, underestimating the size of the 2006 year class at age 1 . The 2007 to 2010 and 2012 year classes also seem to have been underestimated at age 1. The IBTS-Q3 survey shows low mortality for the 2006 year class, and a potential under estimate of the 2007 year class at age 1 ; however, numbers at age 2 in the 2007 yearclass may well
be an overestimate. There does not appear to be a problem estimating age 1 in the 2008 or subsequent year classes in the IBTS-Q3 survey.
The consistency within surveys is assessed using correlation plots in Figures 12.11 and 12.12. These indicate that the IBTS Q1 and Q3 surveys both show good internal consistency across ages. The log CPUE plots by survey (Figure 12.13) support the conclusion of good internal consistency.
Figures 12.14 to 12.16 summarize the results of a SURBAR analysis using the available whiting surveys. These show a well-specified analysis in which the data agree broadly with the separability assumptions in the model and uncertainty bounds are fairly tight.

### 12.1.6.2 Exploratory catch-at-age-based analyses

Catch curves for the catch data are plotted in Figure 12.17 and show numbers-at-age on the $\log$ scale linked by cohort. This shows partial recruitment to the fishery up to age 2 for some cohorts. Also evident is the persistence of the 1999 to 2001 year classes in past catches and the recent low catches of the 2002-2010 year classes.

The negative gradients of log catches per cohort, averaged over ages 2-6 and interpreted here as a rough proxy for fishing mortality over those ages, are given in Figure 12.18. The gradients (since the 2002 year class) appear to be fluctuating around a mean level that is lower than the mean level before the 1998 year-class, which suggests that recent fishing mortality is likely to be lower than in the past.
Within cohort correlations between ages are presented in Figure 12.19. In general, catch numbers correlate well between cohorts with the relationship breaking down as cohorts are compared across increasing age gaps.

Single fleet XSA runs were conducted to compare trends in the catch data with using survey data for quarter 1 and 3 separately. These used the same procedure as this year's final assessment. Summary plots of these runs are presented in Figure 12.20. The population trends from each survey are consistent; however, the mean F estimates differ considerably throughout the time-series. In recent years estimates in SSB, fishing mortality and recruitment have been similar. Residual patterns (Figure 12.21) show that the 2006 year class has a large negative residual at age 1 for both surveys (and particularly IBTS Q1). In quarter 1, residuals for age 1 have been larger in some years $(2006,2013)$.

### 12.1.6.3 Conclusions drawn from exploratory analyses

Catch curve analysis and correlation plots show that in general both surveys and catch data track cohorts well and are internally consistent. However, beginning with the 2006 year class, the IBTS Q1 appears to be underestimating the abundance of age 1 and 2 whiting. In previous assessments, this had implications for the estimation of recruitment at age 1 in 2007 and resulted in a considerable retrospective bias in recruitment.

### 12.1.6.4Final assessment

The final assessment used an FLXSA model fitted to the combined landings, discard and industrial bycatch data for the period 1990-2015. This is the same procedure as last year and that agreed at WKROUND (ICES-WKROUND 2013). The settings are provided in the table below.
Catch-at-age data 1990-2015
Ages 1-8+
Calibration period

Survey: IBTS Q1
1990-2016
Ages 1-5
Survey: IBTS Q3 1991-2015
Ages 1-5
Catchability independent of stock size from
Age 1
Catchability plateau
Weighting
Shrinkage
Shrinkage SE
Age 4
No taper weighting
Last 3 years and 4 ages

Minimum SE for survivors' estimates
2.0
0.3

Diagnostics for the final XSA run are given in Tables 12.15. Residual plots are presented in Figure 12.23. These show that the IBTS Q3 survey fits more closely to the model and the catch data, than the IBTS Q1 survey which demonstrated considerable year effects towards the end of the time series. This indicates that the model is effectively paying less attention to the Q1 survey than to the Q3 survey, and this is borne out by Figure 12.24 which shows the contribution of each tuning fleet to the estimation of survivors in the most recent year.

Finally, Figure 12.22 compares the SURBAR results with the final XSA assessment. The mean Z (total mortality) estimates show year-to-year variation, but the trends in all outputs are very similar.

Fishing mortality estimates are presented in Table 12.16, estimated stock numbers in Table 12.17 and the assessment summary in Table 12.18 and Figure 12.25.

A retrospective analysis is shown in Figure 12.26.1. This shows a consistent bias in recruitment from 2006 to 2010. The largest revision in recruitment is for recruitment in 2008 (the 2007 year class) which coincides with large negative residuals and the flat catch curve in the IBTS-Q1 survey (Figure 12.10). This translates directly to a large revision of SSB in 2008. However, the last four retrospective runs are very consistent for SSB and fishing mortality. This may indicate that previous data problems have been corrected, although it may be too early to say whether the retrospective bias has actually been eliminated. Relative differences are illustrated in Figure 12.26.2, they are calculated as a percentage difference per year from the final year assessment. For each of the retrospective percentage difference plots, the terminal values of the past 7 years are used to calculate Mohn's rho, $(\varrho(S S B)=-0.026, ~ \varrho(F)=0.026, \varrho(\operatorname{Rec})=-0.01)$.

### 12.1.7 Historical stock trends

Historical trends for catch, mean F, SSB and recruitment are presented in Figure 12.25. These show that mean F has been declining and has reached the minimum of the post1990 time-series in 2012, but is increasing in the recent years. The SSB has decreased after recent increases; and recruitment is fluctuating around a recent low average. In the most recent year, landings, discards and industrial bycatch have also all remained at or around a recent average. The stock-recruitment plot in Figure 12.27 shows some evidence of a weak positive relationship between SSB and subsequent recruitment, although such evidence is not compelling.

### 12.1.8 Biological reference points

Due to the shape of the yield per recruit (YPR) curve, a maximum is often not reached, and $F_{\text {max }}$ has therefore not been defined for several years. The WG considers that YPR $F$ reference points are not applicable to this stock since $F_{\text {max }}$ is undefined in most years, and the estimate of $\mathrm{F}_{0.1}$ is very variable in recent years (see ICES-WGNSSK, 2009). A long-term average selection pattern could be used to stabilize $\mathrm{F}_{0.1}$ or a long term average of $\mathrm{F}_{0.1}$ could be interpreted as a sensible reference point. The 2013 benchmark meeting (ICES-WKROUND 2013) attempted to calculate Fmsy for North Sea whiting, but concluded that this value was inestimable using standard equilibrium considerations and would need to be determined as part of a management strategy evaluation.

After the considerable revisions in the 2012 assessment, caused by new estimates of natural mortality, the target $F$ of 0.3 was no longer considered applicable. The management plan was re-evaluated in October 2013 (ICES 2013) and ICES advised that updating the target F from 0.3 to 0.15 within the management plan.

New revisions of natural mortalities were presented at WGSAM 2014. The new natural mortality values from the 2014/2015 key run are used in the current assessment. Due to the new natural mortalities, the recruitment estimates and SSB decreased in the assessment, an Interbenchmark was performed for whiting in the North Sea and Division 7.d (ICES 2016). This included a comparison of assessment results, Eqsim runs and MSE. On the basis of the 2015 assessment using the new natural mortalities the target F of 0.15 leads to maximum probabilities above $5 \%$ of SSB falling below $\mathrm{B}_{\mathrm{lim}}$, which is considered precautionary. This is under the assumption that recruitment stays within a medium-low range. Therefore, a target F of 0.15 together with a TAC constraint of $15 \%$ according to the EU-Norway Management Plan may not be sufficient to keep SSB above $B_{l i m}$. It was concluded to use an MSY approach with FMSY of 0.15 and an additional check of SSB relative to Blim. The target fishing mortality can then be adapted at very low biomass levels. Until additional information becomes available, it is considered that the lowest observed SSB (SSB in 2007, 172741 t in the 2015 assessment) can be used as a $\mathrm{B}_{\mathrm{lim}}$ reference point. As a result new reference points are listed in Table 12.19.

### 12.1.9 Recruitment estimates

RCT3 input data are presented in Table 12.20, and RCT3 output is presented in Table 12.21. The RCT3 estimate of recruitment at age 1 in 2016 (that is, the 2015 year-class) was 2900 million. Following the approach taken last year, and subsequently formalized in the benchmark report (ICES-WKROUND 2013), the WG agreed to use the RCT3 estimates for recruitment in 2016, and the long-term geometric mean for recruitment in 2017 and beyond in the short-term forecast. The geometric mean of all recruitments excluding the most recent year is 2443 million (Table 12.22).

### 12.1.10 Short-term forecasts

A short-term forecast was carried out based on the final FLXSA assessment. FLXSA survivors from 2015 were used as input population numbers for ages 2 and older in 2016. Recruitment assumptions are detailed in the preceding section.

The exploitation pattern was chosen as the mean exploitation pattern over the years 2013-2015. A simple mean F would have led to bias in forecast F, given the recent changes in $\mathrm{F}_{2-6}$, so this exploitation pattern was scaled to the mean $\mathrm{F}_{2-6}$ in 2015 for forecasts. Partial F at age for each catch component was estimated by splitting the forecast F at age using the mean proportion in the catch of each catch component over the
years 2013-2015. The F at age used in the forecast is compared with the F at age estimates for 2013-2015 in Figure 12.27.

Mean weights at age are generally consistent over the recent period but there are trends at several ages (Figure 12.5). To avoid introducing bias, therefore, the 2015 estimates were used for the purposes of forecasting.

The inputs to the short-term forecast are given in Tables 12.22 to 12.23 , and results are presented in Table 12.24. The MFDP program was used to carry out the forecasts, since there is no available function currently within FLR to account for industrial bycatch in forecasts and WGNSSK could not complete the coding required to address this in the time available.

No TAC constraint was applied in the intermediate year since it is not considered that fishing will stop when the TAC is reached.

Assuming mean $\mathrm{F}_{2016}$ to equal mean $\mathrm{F}_{2015}$, results in human consumption landings in 2016 of 18537 t from a total catch of 33601 t , giving an SSB in 2017 of 310363 t (Table 12.24).

Carrying the same fishing mortality forward into 2017 (the status quo F option) would result in landings of 19926 t out of total catches of 35900 t , and would result in an SSB of 317115 t in 2018 (a $2 \%$ increase in SSB relative to 2017). Applying the $\mathrm{F}_{\text {msy }}$ of 0.15 in 2017 would generate landings of 12679 t out of total catches of 23679 t , and result in an SSB of 327559 t in 2018 (a $5.5 \%$ increase in SSB relative to 2017). In 2018, SSB would be above $\mathrm{B}_{\mathrm{lim}}$. F of 0.15 would also cause the TAC to be changed by $-29 \%$.

### 12.1.11 MSY estimation and medium-term forecasts

No medium-term forecasts or MSY estimation were conducted during the WG meeting.

### 12.1.12 Quality of the assessment

Previous meetings of WGNSSK and the benchmark workshop (ICES-WKROUND 2009, ICES-WKROUND 2013) have concluded that the historical survey data and commercial catch data contain different signals concerning the stock. Analyses by Working Group members and by the ICES Study Group on Stock Identity and Management Units of Whiting (ICES-SGSIMUW 2005) indicate that data since the early- to mid1990s are sufficiently consistent to undertake a catch-at-age analysis calibrated against survey data from 1990. This has been taken forward into prediction for catch option purposes. However, due to the lack of concordance in the data pre-dating the early 1990s, WGNSSK considers that it is not possible categorically to classify the current state of the stock with reference to precautionary reference points as the biomass reference points are derived from a consideration of the stock dynamics dating from a time when the commercial catch-at-age data and the survey data conflict. Precautionary reference points must be reconsidered following the ongoing management strategy evaluation.

The IBTS-Q1 survey is showing a step change in catchability of young fish (especially age 1). The reason for this is unclear, but it appears to have happened after the 2006 survey. This represents a model misspecification, as the current model (FLXSA) assumes constant catchability through time.
Due to the likely population structuring in the North Sea and Eastern Channel, it is probable that the overall stock estimates may not reflect trends in more localized areas.

Given the spatial structure of the whiting stock and of the fleets exploiting it, it is important to have data that covers all fleets. Considering that age 1 and age 2 whiting make up a large proportion of the total stock biomass, good information of the discarding practices of the major fleets is important.
The survey information for Division 7. d were not available in a form that could be used by WGNSSK. Due to the recent changes in distribution of the stock, tuning information from this area would be extremely useful, and could improve the estimate of recruitment in the most recent year. However, previous analyses of the survey in Division 7.d showed it did not track cohorts well (ICES-WKROUND 2009).

Age distributions and mean weights at age have been estimated for the industrial bycatch from 2006 to 2010. This was due to low sampling levels of the Danish industrial bycatch fisheries. In recent years no samples of industrial bycatch were available. Age distributions and weights at age were inferred from sampling of landings and discards from other fleets.

There have been issues with regard to the whiting age reading of Norwegian samples. The readings did not correlate well with readings from the Netherlands. The issue mainly affects age readings for the IBTS survey. Until the issue is resolved, the IBTS survey data is used as in previous years.

The historical performance of the assessment is summarized in Figure 12.29. The difference in recruitment estimates in the current assessment is caused by the new natural mortality estimates.

### 12.1.13 Status of the stock

For North Sea whiting, SSB has a generally downwards trend since the start of the assessment time-series. SSB is estimated to be well above Blim since 2008 (Figure 12.24). The stock, at the level of the entire North Sea and Eastern Channel, was at an historical low level during 2005 to 2008 (relative to the period since 1990), and the recent increase in SSB is in large part due to relatively improved perception of recruitment in 2008 and 2011. All indications are that fishing mortality has been declining over most of the timeseries, currently fluctuating around a low level with a small increase in recent years. The level of recruitment has been generally low since 2003, with recruitment since 2014 above the average of the recent years. Recruitment is varying around a recent mean, but that mean is low relative to the rest of the time series and whiting biomass is likely to decline in future (even at low fishing mortality rates) until the appearance of the next good year class.

### 12.1.14 Management considerations

In 1997, 2003 to 2007 and 2012 to 2013, the whiting stock produced the lowest recruitments in the series. Whiting recruitment (estimated largely from the IBTS-Q1 and IBTSQ3 surveys) was underestimated substantially in 2007 and 2008 resulting in low forecasts of recruitment and recommendations of reduced TACs due to the perception of critically low recruitment. Subsequent recruitment is below the long-term average.
Whiting mature at age 2 and grow quickly at young ages; therefore an increase in SSB is seen the year immediately after a good recruitment. Managers should consider the age structure of the population as well as the SSB since at low stock sizes short term forecasts are highly sensitive to recruitment assumptions.
Catches of whiting have been declining since 1980 (from 243570 t in 1979 to 25078 t in 2012, including discards and industrial bycatch). Catch rates from localized fleets may
not represent trends in the overall North Sea and English Channel population. The localized distribution of the population is known to be resulting in substantial differences in the quota uptake rate. This is likely to result in localized discarding problems that should be monitored carefully
Whiting are caught in mixed demersal roundfish fisheries, fisheries targeting flatfish, the Nephrops fisheries, and the Norway pout fishery. The current minimum mesh-size in the targeted demersal roundfish fishery in the northern North Sea has resulted in reduced discards from that sector compared with the historical discard rates. Mortality may have increased on younger ages due to increased discarding in recent years as a result of recent changes in fleet dynamics of Nephrops fleets and small mesh fisheries in the southern North Sea. The bycatch of whiting in the Norway pout and sandeel fisheries is dependent on activity in that fishery, which has recently declined after strong reductions in the fisheries. Industrial bycatches are considered low in the forecast. A larger catch allocation for bycatch may be required if industrial effort increases.

Catches of whiting in the North Sea are also likely to be affected by the effort reduction seen in the targeted demersal roundfish and flatfish fisheries, although this will in part be offset by increases in the number of vessels switching to small mesh fisheries. It is important to consider both the species-specific assessments of these species for effective management, but also the broader mixed-fisheries context. This is not straightforward when stocks are managed via a series of single-species management plans that do not incorporate such mixed-stocks considerations. WGMIXFISH monitors the consistency of the various single-species management plans and TAC advice under current effort schemes, in order to estimate the potential risks of quota over- and under shooting for the different stocks, and it was demonstrated that the current basis for whiting advice was not consistent with other single-stock management objectives. It is recommended that the ongoing discussions about the whiting management plan takes into account such mixed-fisheries considerations before implementation.
Recent measures to improve survival of young cod, such as the Scottish Credit Conservation Scheme, and increased uptake of more selective gear in the North Sea and Skagerrak, should be encouraged for whiting. There is a mismatch between quota allocations, derived from relative stability criteria, and the access of the various fisheries to the resource which has changed because of changes in the distribution of whiting.

ICES has developed a generic approach to evaluate whether new survey in-formation that becomes available in September forms a basis to update the advice. ICES will publish new advice in November 2016 if this is the case for this year.

### 12.1.15 Frequency of assessment

Stocks are considered candidates for biennial assessment if:

- The advice for the stock has been 0-catch or equivalent for the latest three advice years.


## Does not apply for North Sea whiting

Stocks are considered candidates for biennial assessment if the following criteria are fulfilled simultaneously:

- Life span (i.e. maximum normal age) of the species is larger than 5 years

The stock status in relation to the reference points is according to the MSY criteria F (latest assessment year) $<=1.1 \times \mathrm{F}_{\text {msy }}$ OR if $\mathrm{F}_{\text {msy }}$ range has been defined: F (latest assessment year) is $<=F_{\text {upper }}$ (upper bound in $F$ range)
NO, $F(2015)=0.227 F_{\text {msу }} \times 1.1=0.165$
AND SSB(start of intermediate year) $>=$ MSY $B_{\text {trigger }}$
YES, SSB(2015) = $246870>=241837$

- The average contribution to the catch in numbers of the recruiting year class in latest 5 years is less than $25 \%$ of the total catch in numbers. Should be calculated as the average over the latest five years of the catch in numbers of first age divided by the total catch in number by year.

NO, the average contribution of recruits (age 1 ) in numbers in the last 5 years is $\mathbf{6 5 \%}$.

- The retrospective pattern, based on a seven years peel of Mohn's Rho index, shows that F is consistently underestimated by less than $20 \%$

YES, $\rho=0.026<0.2$
The formula to be used in the calculations is:
$\rho=\frac{1}{7} \sum_{u=Y-7}^{Y-1}\left(1-\frac{F_{u, u}}{F_{u, Y}}\right)$. The result should be $<0.20$,
where $F_{u, u}$ is F in year u estimated from an assessment that ends in year u , and $F_{u, Y}$ is the F in year u estimated from the most recent assessment (which ends in year Y )

Results of the frequency of assessment analysis are summarized in Table 12.24.

## In conclusion, North Sea whiting does not qualify for a change in assessment fre-

 quency.
### 12.1.16 References

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Table 12.1: Whiting in Subarea 4 and Division 7.d. Whiting in Subarea 4. Nominal landings (in tonnes) as officially reported to ICES, WG estimates of catch components, and TACs.

| Year | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| belgium. 4 | 1040 | 913 | 1030 | 944 | 1042 | 880 | 843 | 391 | 268 | 529 | 536 | 454 | 270 |
| denmark. 4 | 1206 | 1528 | 1377 | 1418 | 549 | 368 | 189 | 103 | 46 | 58 | 105 | 105 | 96 |
| france. 4 | 4951 | 5188 | 5115 | 5502 | 4735 | 5963 | 4704 | 3526 | 1908 | NA | 2527 | 3455 | 3314 |
| germany. 4 | 692 | 865 | 511 | 441 | 239 | 124 | 187 | 196 | 103 | 176 | 424 | 402 | 354 |
| netherlands. 4 | 3273 | 4028 | 5390 | 4799 | 3864 | 3640 | 3388 | 2539 | 1941 | 1795 | 1884 | 2478 | 2425 |
| norway. 4 | 55 | 103 | 232 | 130 | 79 | 115 | 66 | 75 | 65 | 68 | 33 | 44 | 47 |
| sweden. 4 | 16 | 48 | 22 | 18 | 10 | 1 | 1 | 1 | 0 | 9 | 4 | 6 | 7 |
| england.wales. 4 | 2338 | 2676 | 2528 | 2774 | 2722 | 2477 | 2329 | 2638 | 2909 | 2268 | 1782 | 1301 | 1322 |
| scotland. 4 | 23486 | 31257 | 30821 | 31268 | 28974 | 27811 | 23409 | 22098 | 16696 | 17206 | 17158 | 10589 | 7756 |
| uk. 4 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| total.landings. 4 | 41057 | 46606 | 47026 | 47295 | 42214 | 41379 | 35116 | 31567 | 23936 | NA | 24453 | 18834 | 15591 |
| unallocated.landings. 4 | -1123 | 396 | 1816 | 685 | 344 | 829 | -434 | 627 | 246 | NA | 173 | -426 | 721 |
| wg.landings. 4 | 42180 | 46210 | 45210 | 46610 | 41870 | 40550 | 35550 | 30940 | 23690 | 25700 | 24280 | 19260 | 14870 |
| wg.discards. 4 | 52270 | 30840 | 28470 | 41400 | 31840 | 28940 | 27130 | 16660 | 12480 | 22110 | 21931 | 16130 | 17144 |
| wg.ibc. 4 | 51337 | 39755 | 25045 | 20723 | 17473 | 27379 | 5116 | 6213 | 3494 | 5038 | 9160 | 940 | 7270 |
| wg.catch. 4 | 145787 | 116805 | 98725 | 108733 | 91183 | 96869 | 67796 | 53813 | 39664 | 52848 | 55371 | 36330 | 39284 |
| tac.4.2.a | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | 30000 | 29700 | 41000 |


| Year | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| belgium. 4 | 248 | 144 | 105 | 93 | 45 | 115 | 162 | 147 | 74 | 45 | 33 | 46 | 69 |
| denmark. 4 | 89 | 62 | 57 | 251 | 78 | 42 | 79 | 156 | 135 | 131 | 124 | 160 | 215 |
| france. 4 | 2675 | 1721 | 1261 | 2711 | 3336 | 3076 | 2305 | 2644 | 2794 | 1925 | 942 | 1887 | 1130 |
| germany. 4 | 334 | 296 | 149 | 252 | 76 | 76 | 124 | 156 | 111 | 25 | 44 | 31 | 73 |
| netherlands. 4 | 1442 | 977 | 805 | 702 | 618 | 656 | 718 | 614 | 514 | 471 | 495 | 466 | 548 |
| norway. 4 | 39 | 23 | 16 | 17 | 11 | 92 | 73 | 118 | 28 | 94 | 560 | 916 | 1088 |
| sweden. 4 | 10 | 2 | 0 | 2 | 1 | 2 | 4 | 8 | 6 | 4 | 1 | 2 | 5 |
| england.wales. 4 | 680 | 1209 | 2560 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| scotland. 4 | 5734 | 5057 | 3441 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| uk. 4 | NA | NA | NA | 11632 | 12110 | 10391 | 8853 | 7845 | 8892 | 9893 | 11162 | 10248 | 9970 |
| total.landings. 4 | 11251 | 9491 | 8394 | 15660 | 16275 | 14451 | 12318 | 11690 | 12554 | 12588 | 13361 | 13756 | 13098 |
| unallocated.landings. 4 | 801 | 541 | -2286 | 563 | 609 | 972 | 544 | -591 | -751 | -341 | -2023 | -1860 | -510 |
| wg.landings. 4 | 10450 | 8950 | 10680 | 15097 | 15666 | 13479 | 11774 | 12281 | 13305 | 12929 | 15384 | 15616 | 13608 |
| wg.discards. 4 | 26135 | 18142 | 10300 | 14018 | 5206 | 8356 | 5223 | 7853 | 8180 | 5929 | 4198 | 8326 | 10468 |
| wg.ibc. 4 | 2730 | 1210 | 890 | 2190 | 1240 | 0 | 1020 | 1350 | 1750 | 78 | 1530 | 1479 | 2053 |
| wg.catch. 4 | 39315 | 28302 | 21870 | 31305 | 22112 | 21835 | 18017 | 21484 | 23235 | 18936 | 21119 | 25421 | 24076 |
| tac.4.2.a | 16000 | 16000 | 28500 | 23800 | 23800 | 17850 | 15173 | 12897 | 14832 | 17056 | 18932 | 16092 | 13678 |

Table 12.2: Whiting in Subarea 4 and Division 7.d. Whiting in Division 7.d. Nominal landings (in tonnes) as officially reported to ICES, WG estimates of catch components, and TACs.

| Year | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| belgium.7.d | 83 | 83 | 66 | 74 | 61 | 68 | 84 | 98 | 53 | 48 | 65 | 75 | 58 |
| france.7.d | NA | NA | 5414 | 5032 | 6734 | 5202 | 4771 | 4532 | 4495 | NA | 5875 | 6338 | 5172 |
| netherlands.7.d | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 32 | 6 | 14 | 67 | 19 |
| england.wales.7.d | 239 | 292 | 419 | 321 | 293 | 280 | 199 | 147 | 185 | 135 | 118 | 134 | 112 |
| scotland.7.d | 0 | 0 | 24 | 2 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| uk.7.d | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| total.landings.7.d | NA | NA | 5923 | 5429 | 7088 | 5551 | 5056 | 4779 | 4765 | NA | 6072 | 6614 | 5361 |
| unallocat.landings.7.d | NA | NA | 203 | 219 | 468 | 161 | 106 | 159 | 165 | NA | 1772 | 814 | -439 |
| wg.landings.7.d | 3480 | 5720 | 5740 | 5210 | 6620 | 5390 | 4950 | 4620 | 4600 | 4430 | 4300 | 5800 | 5800 |
| wg.discards.7.d | 3330 | 4220 | 4090 | 2970 | 3850 | 3240 | 3370 | 3000 | 3210 | 3570 | 4129 | 3109 | 1356 |
| wg.catch.7.d | 6810 | 9940 | 9830 | 8180 | 10470 | 8630 | 8320 | 7620 | 7810 | 8000 | 8429 | 8909 | 7156 |
| tac.7b.k | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | 22000 | 21000 | 31700 |


| Year | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| belgium.7.d | 67 | 46 | 45 | 73 | 75 | 69 | 71 | 88 | 78 | 66 | 95 | 89 | 121 |
| france.7.d | 6654 | 5006 | 4638 | 3487 | 3135 | 2875 | 6248 | 5512 | 4833 | 3093 | 3076 | 2115 | 3065 |
| netherlands.7.d | 175 | 132 | 128 | 117 | 118 | 162 | 112 | 275 | 282 | 437 | 650 | 663 | 558 |
| england.wales.7.d | 109 | 99 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| scotland.7.d | 0 | 0 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| uk.7.d | NA | NA | 90 | 72 | 63 | 87 | 138 | 258 | 271 | 261 | 472 | 345 | 365 |
| total.landings.7.d | 7005 | 5283 | 4901 | 3749 | 3391 | 3193 | 6569 | 6133 | 5464 | 3857 | 4293 | 3212 | 4109 |
| unalloc.landings.7.d | 1295 | 933 | 111 | 306 | 137 | -1278 | -77 | 194 | 400 | -246 | 343 | 82 | 11 |
| wg.landings.7.d | 5710 | 4350 | 4790 | 3443 | 3254 | 4471 | 6646 | 5939 | 5064 | 4103 | 3950 | 3130 | 4098 |
| wg.discards.7.d | 604 | 907 | 2219 | 2291 | 1763 | 1943 | 2477 | 3727 | 3538 | 2446 | 1778 | 2125 | 2960 |
| wg.catch.7.d | 6314 | 5257 | 7009 | 5734 | 5017 | 6414 | 9123 | 9666 | 8602 | 6549 | 5728 | 5255 | 7059 |
| tac.7b.k | 31700 | 27000 | 21600 | 19940 | 19940 | 19940 | 16949 | 14407 | 16568 | 19053 | 24500 | 20668 | 17742 |

Table 12.3.1: Whiting in Subarea 4 and Division 7.d. Description of InterCatch raising procedure using Table 2 of CatchAndSampleData.Tables.txt. SOP.

| CatchCategory | SOP |
| :--- | :--- |
| Discards | 1.002 |
| Landings | 1.032 |

Table 12.3.2: Whiting in Subarea 4 and Division 7.d. Description of InterCatch raising procedure using Table 2 of CatchAndSampleData.Tables.txt. Summary of imported and raised data.

| CatchCategory | RaisedOrImported |  | CATON |
| :--- | :--- | :--- | :--- |
| percent |  |  |  |
| Discards | Raised | 2292643 | 17 |
| Discards | Imported | 11100197 | 83 |
| Landings | Imported | 19396023 | 100 |

Table 12.3.3: Whiting in Subarea 4 and Division 7.d. Description of InterCatch raising procedure using Table 2 of CatchAndSampleData.Tables.txt. Summary of the imported/Raised/SampledOrEstimated data.

|  |  |  | SampledOrEstimated <br> distribution | CATON | perc |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Landings | Imported | Sampled | 13744977 | 69 |  |
| Landings | Imported | Estimated | 6265110 | 31 |  |
| Discards | Imported | Sampled | 6637692 | 49 |  |
| Discards | Imported | Estimated | 4451738 | 33 |  |
| Discards | Raised | Estimated | 2338849 | 17 |  |

Table 12.3.4: Whiting in Subarea 4 and Division 7.d. Description of InterCatch raising procedure using Table 2 of CatchAndSampleData.Tables.txt. Summary of the imported/Raised/SampledOrEstimated data by area.

| CatchCategory | RaisedOrImported | SampledOrEstimated distribution | Area | CATON | percent |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | Imported | Sampled | 7.d | 2676203 | 65 |
| Landings | Imported | Estimated | 7.d | 1421845 | 35 |
| Discards | Imported | Sampled | 7.d | 1631383 | 55 |
| Discards | Raised | Estimated | 7.d | 966053 | 33 |
| Discards | Imported | Estimated | 7.d | 363286 | 12 |
| Landings | Imported | Sampled | 4 | 11068775 | 70 |
| Landings | Imported | Estimated | 4 | 4843266 | 30 |
| Discards | Imported | Sampled | 4 | 5006309 | 48 |
| Discards | Imported | Estimated | 4 | 4088452 | 39 |
| Discards | Raised | Estimated | 4 | 1372796 | 13 |

Table 12.4: Whiting in Subarea 4 and Division 7.d. Total catch numbers at age (thousands). Age 8 is a plus-group. Ages 1-8+ and years 1990-2015 are included in the final assessment.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 687238 | 418910 | 313391 | 242370 | 90047 | 7563 | 7565 | 1851 | 253 | 11 | 9 | 4 | 0 | 0 | 0 | 0 | 277 |
| 1979 | 476383 | 615524 | 467537 | 218283 | 100975 | 29267 | 3111 | 1657 | 264 | 35 | 1 | 4 | 0 | 0 | 0 | 0 | 304 |
| 1980 | 332209 | 265359 | 416008 | 286077 | 90718 | 52969 | 10751 | 1152 | 689 | 58 | 14 | 5 | 1 | 0 | 0 | 0 | 767 |
| 1981 | 516869 | 162899 | 346343 | 266517 | 102295 | 27776 | 12297 | 3540 | 244 | 45 | 37 | 1 | 0 | 0 | 0 | 0 | 326 |
| 1982 | 101058 | 192640 | 114444 | 245246 | 88137 | 26796 | 6909 | 2082 | 400 | 53 | 26 | 4 | 1 | 0 | 0 | 0 | 484 |
| 1983 | 668604 | 205646 | 184746 | 118412 | 131508 | 37231 | 8688 | 1780 | 794 | 101 | 35 | 0 | 0 | 0 | 0 | 0 | 930 |
| 1984 | 157819 | 323408 | 175965 | 124886 | 49505 | 59817 | 13860 | 2964 | 410 | 182 | 21 | 0 | 0 | 0 | 0 | 0 | 613 |
| 1985 | 186723 | 203321 | 141716 | 82037 | 37847 | 14420 | 17445 | 3328 | 805 | 89 | 9 | 1 | 0 | 0 | 0 | 0 | 904 |
| 1986 | 225201 | 576731 | 167077 | 169577 | 46517 | 13367 | 3487 | 3975 | 497 | 71 | 0 | 1 | 0 | 0 | 0 | 0 | 569 |
| 1987 | 84863 | 267051 | 368229 | 122748 | 85240 | 11392 | 4556 | 928 | 929 | 98 | 7 | 0 | 0 | 0 | 0 | 0 | 1035 |
| 1988 | 416924 | 430344 | 307429 | 179502 | 39635 | 17901 | 2175 | 544 | 59 | 72 | 37 | 0 | 0 | 0 | 0 | 0 | 168 |
| 1989 | 87325 | 331672 | 173676 | 191942 | 78464 | 14367 | 5050 | 516 | 291 | 36 | 6 | 1 | 0 | 0 | 0 | 0 | 334 |
| 1990 | 289174 | 258102 | 501372 | 127966 | 84147 | 31102 | 1934 | 719 | 93 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 109 |
| 1991 | 1058000 | 135797 | 194921 | 184960 | 36290 | 25554 | 5339 | 526 | 249 | 17 | 1 | 0 | 0 | 0 | 0 | 0 | 268 |
| 1992 | 259390 | 230302 | 167478 | 87819 | 91081 | 11654 | 6634 | 2546 | 104 | 7 | 1 | 0 | 0 | 0 | 0 | 0 | 112 |
| 1993 | 628301 | 223425 | 172048 | 125599 | 46181 | 45300 | 3899 | 1501 | 682 | 56 | 15 | 0 | 0 | 0 | 0 | 0 | 754 |
| 1994 | 218286 | 191544 | 158369 | 97559 | 51040 | 18683 | 17905 | 1258 | 441 | 73 | 0 | 0 | 0 | 0 | 0 | 0 | 514 |
| 1995 | 1597900 | 148170 | 144023 | 112416 | 35649 | 15062 | 5117 | 4472 | 315 | 101 | 54 | 0 | 0 | 0 | 0 | 0 | 470 |
| 1996 | 96515 | 86318 | 118910 | 99644 | 48303 | 14088 | 4638 | 1281 | 897 | 166 | 24 | 6 | 2 | 0 | 0 | 0 | 1095 |
| 1997 | 19001 | 60946 | 80471 | 84336 | 41975 | 18304 | 3333 | 1012 | 304 | 135 | 16 | 0 | 0 | 0 | 0 | 0 | 456 |
| 1998 | 72289 | 92557 | 50361 | 43423 | 36295 | 17627 | 6343 | 1416 | 306 | 66 | 33 | 0 | 0 | 0 | 0 | 0 | 405 |
| 1999 | 76976 | 189162 | 95416 | 45920 | 33921 | 18271 | 7443 | 2021 | 565 | 95 | 12 | 0 | 0 | 0 | 0 | 0 | 672 |
| 2000 | 1970 | 82545 | 129582 | 63706 | 23913 | 16198 | 8758 | 4309 | 969 | 244 | 47 | 3 | 0 | 0 | 0 | 0 | 1264 |
| 2001 | 18011 | 52567 | 83086 | 52076 | 20799 | 9256 | 4826 | 2233 | 896 | 246 | 124 | 2 | 0 | 0 | 0 | 0 | 1268 |
| 2002 | 135848 | 51338 | 62462 | 84600 | 34659 | 8098 | 2048 | 1461 | 621 | 102 | 13 | 9 | 9 | 0 | 0 | 0 | 755 |
| 2003 | 60744 | 83680 | 111144 | 55866 | 41840 | 14218 | 2358 | 473 | 329 | 50 | 16 | 1 | 0 | 0 | 0 | 0 | 397 |
| 2004 | 34210 | 47967 | 23009 | 32557 | 30401 | 21755 | 8342 | 1351 | 197 | 93 | 12 | 1 | 4 | 0 | 0 | 0 | 307 |
| 2005 | 17621 | 47805 | 34627 | 12204 | 18146 | 14931 | 8979 | 3041 | 540 | 83 | 29 | 1 | 0 | 0 | 0 | 0 | 654 |
| 2006 | 15673 | 73908 | 42198 | 21651 | 8642 | 15077 | 11822 | 4618 | 1300 | 142 | 14 | 0 | 0 | 0 | 0 | 0 | 1457 |
| 2007 | 2490 | 39041 | 34001 | 24900 | 9905 | 4009 | 7657 | 5267 | 2559 | 476 | 82 | 0 | 0 | 0 | 0 | 0 | 3117 |
| 2008 | 5631 | 62164 | 28301 | 22741 | 13571 | 4305 | 1848 | 3954 | 2134 | 631 | 143 | 43 | 0 | 0 | 0 | 0 | 2951 |
| 2009 | 2362 | 19919 | 56301 | 14922 | 11605 | 5331 | 1409 | 613 | 1504 | 942 | 341 | 49 | 1 | 0 | 0 | 0 | 2837 |
| 2010 | 1224 | 26266 | 60426 | 24826 | 8016 | 5394 | 2867 | 518 | 650 | 567 | 239 | 54 | 1 | 0 | 0 | 0 | 1510 |
| 2011 | 612 | 32894 | 59451 | 27509 | 14825 | 3331 | 2179 | 1032 | 119 | 47 | 92 | 55 | 0 | 0 | 0 | 0 | 312 |
| 2012 | 1854 | 28438 | 29366 | 22034 | 17656 | 6541 | 2406 | 1215 | 330 | 86 | 52 | 18 | 55 | 0 | 5 | 0 | 546 |
| 2013 | 4979 | 19972 | 17442 | 30164 | 16063 | 11179 | 3598 | 781 | 366 | 132 | 3 | 0 | 0 | 0 | 0 | 0 | 501 |
| 2014 | 5540 | 43756 | 20633 | 21001 | 21876 | 10837 | 4167 | 1269 | 242 | 119 | 16 | 0 | 0 | 0 | 0 | 0 | 377 |
| 2015 | 3746 | 39951 | 60807 | 14005 | 9152 | 13356 | 4488 | 1423 | 441 | 15 | 29 | 4 | 0 | 0 | 0 | 0 | 489 |

Table 12.5: Whiting in Subarea 4 and Division 7.d. Landings numbers at age (thousands). Age 8 is a plus-group. Ages 1-8+ and years 1990-2015 are included in the final assessment.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | $8+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 0 | 14793 | 99836 | 155424 | 76829 | 6693 | 7202 | 1837 | 253 | 11 | 9 | 4 | 0 | 0 | 0 | 0 | 277 |
| 1979 | 8 | 8488 | 108548 | 144343 | 89093 | 26584 | 3011 | 1617 | 250 | 35 | 1 | 4 | 0 | 0 | 0 | 0 | 290 |
| 1980 | 0 | 3656 | 62405 | 152570 | 68422 | 41430 | 9911 | 1135 | 689 | 58 | 14 | 5 | 1 | 0 | 0 | 0 | 767 |
| 1981 | 6 | 4240 | 69211 | 104348 | 78253 | 23698 | 12036 | 3530 | 244 | 45 | 37 | 1 | 0 | 0 | 0 | 0 | 326 |
| 1982 | 0 | 10890 | 46703 | 124656 | 59393 | 21376 | 5664 | 2058 | 400 | 53 | 26 | 4 | 1 | 0 | 0 | 0 | 484 |
| 1983 | 1 | 10568 | 68640 | 67312 | 101342 | 31266 | 8330 | 1730 | 784 | 101 | 35 | 0 | 0 | 0 | 0 | 0 | 921 |
| 1984 | 0 | 14388 | 62693 | 99204 | 41277 | 51745 | 12735 | 2813 | 410 | 182 | 21 | 0 | 0 | 0 | 0 | 0 | 613 |
| 1985 | 1 | 2288 | 51194 | 57049 | 32340 | 12974 | 16361 | 3238 | 805 | 89 | 9 | 1 | 0 | 0 | 0 | 0 | 904 |
| 1986 | 29 | 12879 | 44500 | 111527 | 37287 | 11285 | 3379 | 3912 | 485 | 71 | 0 | 1 | 0 | 0 | 0 | 0 | 557 |
| 1987 | 22 | 11074 | 72372 | 70504 | 73742 | 10808 | 4506 | 928 | 899 | 98 | 7 | 0 | 0 | 0 | 0 | 0 | 1004 |
| 1988 | 0 | 7462 | 61360 | 94163 | 29147 | 16556 | 2158 | 544 | 56 | 72 | 37 | 0 | 0 | 0 | 0 | 0 | 164 |
| 1989 | 52 | 8636 | 28406 | 77009 | 44307 | 9249 | 3888 | 420 | 208 | 35 | 6 | 1 | 0 | 0 | 0 | 0 | 249 |
| 1990 | 23 | 6910 | 52533 | 43850 | 48537 | 16845 | 1341 | 605 | 91 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 107 |
| 1991 | 410 | 11565 | 42525 | 88974 | 25738 | 21261 | 4581 | 396 | 249 | 17 | 1 | 0 | 0 | 0 | 0 | 0 | 268 |
| 1992 | 298 | 9565 | 44697 | 47843 | 59208 | 9784 | 6099 | 1453 | 99 | 7 | 1 | 0 | 0 | 0 | 0 | 0 | 107 |
| 1993 | 720 | 5957 | 28935 | 63383 | 32819 | 33741 | 2932 | 1339 | 682 | 56 | 15 | 0 | 0 | 0 | 0 | 0 | 753 |
| 1994 | 77 | 17124 | 31351 | 45492 | 36289 | 13920 | 14407 | 914 | 366 | 73 | 0 | 0 | 0 | 0 | 0 | 0 | 439 |
| 1995 | 277 | 8829 | 28027 | 58046 | 27775 | 13652 | 4911 | 4359 | 308 | 101 | 54 | 0 | 0 | 0 | 0 | 0 | 463 |
| 1996 | 1015 | 12517 | 26611 | 47125 | 35828 | 11861 | 4396 | 1103 | 897 | 166 | 24 | 6 | 2 | 0 | 0 | 0 | 1095 |
| 1997 | 608 | 6511 | 23436 | 47717 | 31503 | 15615 | 2931 | 1010 | 289 | 135 | 15 | 0 | 0 | 0 | 0 | 0 | 439 |
| 1998 | 1202 | 17071 | 19828 | 24860 | 24473 | 14579 | 5395 | 1204 | 219 | 64 | 16 | 0 | 0 | 0 | 0 | 0 | 299 |
| 1999 | 68 | 16661 | 26669 | 25504 | 23465 | 14483 | 6554 | 1854 | 514 | 61 | 12 | 0 | 0 | 0 | 0 | 0 | 587 |
| 2000 | 0 | 15384 | 31808 | 28283 | 14241 | 11775 | 6618 | 3758 | 862 | 244 | 47 | 3 | 0 | 0 | 0 | 0 | 1157 |
| 2001 | 150 | 12260 | 28476 | 27293 | 17491 | 8633 | 4503 | 2091 | 877 | 246 | 124 | 2 | 0 | 0 | 0 | 0 | 1249 |
| 2002 | 0 | 2610 | 10346 | 30890 | 22353 | 6712 | 1710 | 1330 | 511 | 99 | 10 | 9 | 9 | 0 | 0 | 0 | 639 |
| 2003 | 20 | 403 | 11613 | 13990 | 18974 | 9513 | 1861 | 443 | 329 | 50 | 16 | 0 | 0 | 0 | 0 | 0 | 396 |
| 2004 | 0 | 3973 | 2812 | 9629 | 13302 | 11846 | 4409 | 747 | 174 | 84 | 12 | 1 | 4 | 0 | 0 | 0 | 274 |
| 2005 | 74 | 11009 | 10414 | 5669 | 10926 | 10283 | 5933 | 2343 | 321 | 78 | 29 | 1 | 0 | 0 | 0 | 0 | 429 |
| 2006 | 11 | 11055 | 11023 | 8494 | 5362 | 12259 | 10161 | 4118 | 1080 | 105 | 6 | 0 | 0 | 0 | 0 | 0 | 1192 |
| 2007 | 140 | 10378 | 14740 | 16491 | 7666 | 3310 | 6681 | 4227 | 2179 | 383 | 77 | 0 | 0 | 0 | 0 | 0 | 2638 |
| 2008 | 0 | 13234 | 12334 | 14120 | 9106 | 3564 | 1519 | 2505 | 1481 | 568 | 143 | 43 | 0 | 0 | 0 | 0 | 2235 |
| 2009 | 2 | 2462 | 31910 | 9615 | 9516 | 4318 | 1252 | 548 | 1156 | 876 | 304 | 49 | 1 | 0 | 0 | 0 | 2386 |
| 2010 | 9 | 3593 | 27147 | 15341 | 4885 | 4063 | 1746 | 363 | 391 | 489 | 230 | 54 | 1 | 0 | 0 | 0 | 1165 |
| 2011 | 0 | 4679 | 22858 | 14952 | 10821 | 2333 | 1484 | 729 | 114 | 42 | 76 | 48 | 0 | 0 | 0 | 0 | 280 |
| 2012 | 213 | 4872 | 13111 | 13014 | 11490 | 4726 | 1590 | 860 | 247 | 76 | 28 | 13 | 49 | 0 | 4 | 0 | 417 |
| 2013 | 7 | 2596 | 7176 | 17656 | 12699 | 9914 | 3208 | 705 | 328 | 122 | 3 | 0 | 0 | 0 | 0 | 0 | 453 |
| 2014 | 0 | 4594 | 9508 | 12019 | 13943 | 8219 | 3006 | 1009 | 198 | 102 | 15 | 0 | 0 | 0 | 0 | 0 | 314 |
| 2015 | 295 | 5293 | 22924 | 7381 | 5945 | 8494 | 3348 | 1084 | 307 | 13 | 27 | 4 | 0 | 0 | 0 | 0 | 350 |

Table 12.6: Whiting in Subarea 4 and Division 7.d. Discards numbers at age (thousands). Age 8 is a plus-group. Ages 1-8+ and years 1990-2015 are included in the final assessment.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 28587 | 52684 | 114965 | 37682 | 7154 | 255 | 110 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 4577 | 473830 | 126724 | 31601 | 7322 | 1263 | 27 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 3144 | 103203 | 250735 | 88399 | 14135 | 10795 | 786 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 867 | 50407 | 96509 | 57403 | 7313 | 1285 | 149 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 18639 | 53753 | 26922 | 52349 | 18230 | 2972 | 343 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 71016 | 152488 | 85318 | 33325 | 23442 | 4309 | 295 | 25 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 |
| 1984 | 16724 | 200589 | 82563 | 16814 | 4437 | 4495 | 1034 | 151 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 8497 | 154232 | 48791 | 15117 | 2985 | 761 | 801 | 65 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 7966 | 404604 | 120492 | 43479 | 5242 | 627 | 108 | 63 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 |
| 1987 | 9978 | 158531 | 202154 | 34824 | 9776 | 582 | 49 | 0 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 31 |
| 1988 | 21321 | 65021 | 87197 | 51135 | 5877 | 846 | 16 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 1989 | 6898 | 150598 | 36712 | 61442 | 21267 | 3276 | 103 | 8 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 |
| 1990 | 147764 | 83152 | 241924 | 33084 | 23009 | 11665 | 246 | 85 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 7208 | 81678 | 82053 | 75035 | 5176 | 1885 | 91 | 60 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 7587 | 105838 | 63830 | 27659 | 23115 | 1231 | 355 | 1064 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 1993 | 48873 | 128248 | 104844 | 51054 | 9205 | 10727 | 521 | 131 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 8352 | 96890 | 102020 | 37751 | 9867 | 2885 | 2338 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 33363 | 53830 | 81783 | 50019 | 7136 | 1336 | 206 | 113 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 |
| 1996 | 4575 | 43126 | 86878 | 49817 | 11506 | 2205 | 240 | 179 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 11525 | 26188 | 34948 | 32473 | 9398 | 2412 | 400 | 2 | 16 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 17 |
| 1998 | 6098 | 50703 | 24200 | 17053 | 11076 | 2987 | 936 | 213 | 87 | 2 | 18 | 0 | 0 | 0 | 0 | 0 | 106 |
| 1999 | 14762 | 96413 | 56365 | 15228 | 9016 | 3104 | 862 | 167 | 51 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 85 |
| 2000 | 1682 | 48162 | 81086 | 24082 | 3075 | 2311 | 1560 | 478 | 107 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 107 |
| 2001 | 17352 | 39826 | 52156 | 23055 | 2795 | 471 | 283 | 142 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 19 |
| 2002 | 1158 | 10597 | 33371 | 45125 | 10136 | 1182 | 218 | 131 | 110 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 116 |
| 2003 | 3584 | 65829 | 94497 | 39301 | 21654 | 4314 | 449 | 30 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
| 2004 | 10478 | 31169 | 15698 | 21879 | 16951 | 9909 | 3922 | 605 | 24 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 33 |
| 2005 | 5499 | 25753 | 23486 | 6041 | 7192 | 4616 | 2992 | 688 | 211 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 216 |
| 2006 | 15662 | 51961 | 25906 | 10935 | 2474 | 2595 | 1598 | 493 | 219 | 37 | 8 | 0 | 0 | 0 | 0 | 0 | 265 |
| 2007 | 2350 | 22508 | 16283 | 7153 | 1784 | 572 | 940 | 1037 | 380 | 93 | 5 | 0 | 0 | 0 | 0 | 0 | 478 |
| 2008 | 5631 | 48929 | 15967 | 8621 | 4465 | 741 | 328 | 1449 | 653 | 63 | 0 | 0 | 0 | 0 | 0 | 0 | 716 |
| 2009 | 2360 | 12411 | 21950 | 4277 | 1715 | 910 | 128 | 62 | 347 | 66 | 37 | 0 | 0 | 0 | 0 | 0 | 450 |
| 2010 | 1215 | 15988 | 30046 | 8121 | 2637 | 1194 | 1082 | 151 | 258 | 77 | 9 | 0 | 0 | 0 | 0 | 0 | 344 |
| 2011 | 612 | 28024 | 34431 | 11770 | 3314 | 866 | 641 | 274 | 5 | 3 | 9 | 5 | 0 | 0 | 0 | 0 | 22 |
| 2012 | 1635 | 23479 | 16165 | 8953 | 6112 | 1796 | 809 | 352 | 82 | 10 | 23 | 5 | 6 | 0 | 0 | 0 | 128 |
| 2013 | 4972 | 17154 | 9653 | 10997 | 2277 | 417 | 116 | 15 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 |
| 2014 | 5540 | 38754 | 10281 | 7915 | 6697 | 1889 | 895 | 171 | 26 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 35 |
| 2015 | 3280 | 32739 | 34337 | 5720 | 2568 | 3938 | 808 | 232 | 102 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 103 |

Table 12.7: Whiting in Subarea 4 and Division 7.d. Industrial bycatch numbers at age (thousands). Age 8 is a plus-group. Ages 1-8+ and years 1990-2015 are included in the final assessment.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 658651 | 351432 | 98590 | 49263 | 6064 | 616 | 252 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 471798 | 133206 | 232266 | 42339 | 4561 | 1420 | 73 | 33 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 |
| 1980 | 329065 | 158500 | 102869 | 45108 | 8162 | 744 | 55 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 515996 | 108252 | 180623 | 104766 | 16729 | 2793 | 112 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 82418 | 127998 | 40818 | 68242 | 10514 | 2448 | 902 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 597587 | 42591 | 30789 | 17774 | 6723 | 1656 | 63 | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 141095 | 108431 | 30709 | 8868 | 3790 | 3577 | 91 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 178224 | 46801 | 41731 | 9871 | 2522 | 685 | 284 | 26 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 217207 | 159249 | 2086 | 14572 | 3987 | 1456 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 74863 | 97446 | 93704 | 17420 | 1722 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 395603 | 357861 | 158872 | 34205 | 4611 | 500 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 80375 | 172438 | 108558 | 53491 | 12890 | 1842 | 1060 | 89 | 71 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 72 |
| 1990 | 141387 | 168040 | 206916 | 51033 | 12601 | 2592 | 346 | 29 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 1991 | 1050381 | 42554 | 70343 | 20951 | 5376 | 2408 | 667 | 70 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 251505 | 114899 | 58952 | 12318 | 8758 | 639 | 180 | 29 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 1993 | 578708 | 89219 | 38270 | 11162 | 4157 | 832 | 445 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 209858 | 77530 | 24998 | 14316 | 4885 | 1878 | 1160 | 337 | 75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 75 |
| 1995 | 1564260 | 85510 | 34213 | 4351 | 738 | 73 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 90925 | 30675 | 5421 | 2702 | 970 | 21 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 6868 | 28247 | 22087 | 4146 | 1074 | 276 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 64989 | 24782 | 6334 | 1511 | 746 | 62 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 62145 | 76088 | 12381 | 5188 | 1440 | 684 | 27 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 288 | 19000 | 16688 | 11341 | 6597 | 2113 | 580 | 73 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 510 | 481 | 2453 | 1728 | 514 | 152 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 134690 | 38131 | 18745 | 8585 | 2170 | 205 | 120 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 57140 | 17448 | 5034 | 2575 | 1213 | 390 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 23732 | 12824 | 4499 | 1049 | 147 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 12049 | 11043 | 726 | 494 | 28 | 32 | 54 | 10 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 |
| 2006 | 0 | 10892 | 5270 | 2222 | 806 | 223 | 63 | 7 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2007 | 0 | 6155 | 2978 | 1256 | 456 | 126 | 36 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 5046 | 2441 | 1030 | 374 | 103 | 29 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2010 | 0 | 6685 | 3234 | 1364 | 495 | 137 | 39 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2011 | 0 | 191 | 2162 | 787 | 691 | 132 | 54 | 30 | 0 | 1 | 7 | 2 | 0 | 0 | 0 | 0 | 11 |
| 2012 | 6 | 87 | 90 | 67 | 54 | 20 | 7 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 2013 | 1 | 222 | 614 | 1511 | 1087 | 848 | 275 | 60 | 28 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 39 |
| 2014 | 0 | 407 | 843 | 1066 | 1237 | 729 | 267 | 89 | 18 | 9 | 1 | 0 | 0 | 0 | 0 | 0 | 28 |
| 2015 | 172 | 1919 | 3546 | 905 | 638 | 924 | 331 | 106 | 32 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 35 |

Table 12.8: Whiting in Subarea 4 and Division 7.d. Total catch mean weights at age (kg). Age 8 is a plus-group. These estimates are also used as stock mean weights at age. Ages 1-8+ and years 19902015 are included in the final assessment.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 0.010 | 0.074 | 0.182 | 0.234 | 0.322 | 0.427 | 0.428 | 0.466 | 0.615 | 0.702 | 1.539 | 0.589 | 0.000 | 0.000 | 0.000 | 0.000 | 0.649 |
| 1979 | 0.009 | 0.098 | 0.166 | 0.259 | 0.301 | 0.411 | 0.455 | 0.492 | 0.578 | 0.617 | 0.737 | 0.515 | 0.000 | 0.000 | 0.000 | 0.000 | 0.582 |
| 1980 | 0.013 | 0.075 | 0.176 | 0.252 | 0.328 | 0.337 | 0.458 | 0.458 | 0.568 | 0.539 | 0.790 | 0.688 | 1.711 | 0.000 | 0.000 | 0.000 | 0.572 |
| 1981 | 0.011 | 0.083 | 0.168 | 0.242 | 0.321 | 0.379 | 0.411 | 0.444 | 0.651 | 0.833 | 1.041 | 0.695 | 0.000 | 0.000 | 0.000 | 0.000 | 0.720 |
| 1982 | 0.029 | 0.061 | 0.184 | 0.253 | 0.314 | 0.376 | 0.478 | 0.504 | 0.702 | 0.772 | 1.141 | 0.853 | 1.081 | 0.000 | 0.000 | 0.000 | 0.736 |
| 1983 | 0.015 | 0.107 | 0.191 | 0.273 | 0.325 | 0.384 | 0.426 | 0.452 | 0.520 | 0.677 | 0.516 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.537 |
| 1984 | 0.020 | 0.089 | 0.188 | 0.271 | 0.337 | 0.382 | 0.391 | 0.463 | 0.575 | 0.514 | 0.871 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.567 |
| 1985 | 0.014 | 0.094 | 0.192 | 0.284 | 0.332 | 0.402 | 0.435 | 0.494 | 0.426 | 0.507 | 0.852 | 0.976 | 0.000 | 0.000 | 0.000 | 0.000 | 0.438 |
| 1986 | 0.015 | 0.105 | 0.183 | 0.255 | 0.318 | 0.378 | 0.475 | 0.468 | 0.540 | 1.226 | 0.990 | 0.535 | 0.000 | 0.000 | 0.000 | 0.000 | 0.626 |
| 1987 | 0.013 | 0.077 | 0.148 | 0.247 | 0.297 | 0.375 | 0.379 | 0.542 | 0.555 | 0.857 | 0.603 | 1.193 | 0.000 | 0.000 | 0.000 | 0.000 | 0.584 |
| 1988 | 0.013 | 0.054 | 0.146 | 0.223 | 0.301 | 0.346 | 0.423 | 0.506 | 0.854 | 0.585 | 0.648 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.694 |
| 1989 | 0.023 | 0.070 | 0.157 | 0.225 | 0.267 | 0.318 | 0.391 | 0.431 | 0.369 | 0.517 | 0.857 | 0.609 | 0.000 | 0.000 | 0.000 | 0.000 | 0.394 |
| 1990 | 0.016 | 0.084 | 0.137 | 0.210 | 0.252 | 0.279 | 0.411 | 0.498 | 0.636 | 0.351 | 0.918 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.594 |
| 1991 | 0.018 | 0.104 | 0.168 | 0.217 | 0.289 | 0.30 | 0.339 | 0.365 | 0.385 | 0.589 | 0.996 | 2.756 | 0.000 | 0.000 | 0.000 | 0.000 | 0.401 |
| 1992 | 0.013 | 0.085 | 0.185 | 0.257 | 0.277 | 0.331 | 0.346 | 0.313 | 0.480 | 0.763 | 1.728 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.506 |
| 1993 | 0.012 | 0.07 | 0.17 | 0.250 | 0.316 | 0.32 | 0.346 | 0.400 | 0.376 | 0.41 | 0.359 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.379 |
| 1994 | 0.013 | 0.084 | 0.167 | 0.255 | 0.328 | 0.382 | 0.376 | 0.419 | 0.438 | 0.392 | 0.499 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.431 |
| 1995 | 0.010 | 0.089 | 0.180 | 0.257 | 0.340 | 0.384 | 0.429 | 0.434 | 0.445 | 0.346 | 0.406 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.419 |
| 1996 | 0.018 | 0.094 | 0.167 | 0.235 | 0.302 | 0.388 | 0.407 | 0.431 | 0.439 | 0.404 | 0.376 | 0.398 | 0.287 | 0.000 | 0.000 | 0.000 | 0.432 |
| 1997 | 0.028 | 0.096 | 0.178 | 0.242 | 0.295 | 0.334 | 0.384 | 0.387 | 0.394 | 0.479 | 0.458 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.422 |
| 1998 | 0.018 | 0.090 | 0.179 | 0.236 | 0.281 | 0.314 | 0.340 | 0.333 | 0.335 | 0.495 | 0.433 | 0.600 | 0.000 | 0.000 | 0.000 | 0.000 | 0.369 |
| 1999 | 0.023 | 0.078 | 0.174 | 0.232 | 0.256 | 0.289 | 0.305 | 0.311 | 0.286 | 0.316 | 0.344 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.292 |
| 2000 | 0.034 | 0.117 | 0.182 | 0.238 | 0.287 | 0.286 | 0.276 | 0.275 | 0.268 | 0.264 | 0.280 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.268 |
| 2001 | 0.024 | 0.101 | 0.192 | 0.244 | 0.282 | 0.267 | 0.298 | 0.284 | 0.286 | 0.301 | 0.315 | 0.505 | 0.000 | 0.000 | 0.000 | 0.000 | 0.292 |
| 2002 | 0.010 | 0.069 | 0.155 | 0.218 | 0.273 | 0.303 | 0.350 | 0.343 | 0.327 | 0.412 | 0.288 | 0.231 | 0.304 | 0.643 | 0.000 | 0.000 | 0.336 |
| 2003 | 0.012 | 0.057 | 0.118 | 0.193 | 0.259 | 0.299 | 0.354 | 0.385 | 0.342 | 0.462 | 0.620 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.368 |
| 2004 | 0.031 | 0.111 | 0.150 | 0.213 | 0.253 | 0.286 | 0.285 | 0.286 | 0.347 | 0.351 | 0.352 | 1.463 | 0.337 | 0.000 | 0.000 | 0.000 | 0.350 |
| 2005 | 0.032 | 0.124 | 0.199 | 0.239 | 0.250 | 0.282 | 0.305 | 0.298 | 0.271 | 0.376 | 0.316 | 0.337 | 0.670 | 0.000 | 0.000 | 0.000 | 0.286 |
| 2006 | 0.093 | 0.131 | 0.180 | 0.231 | 0.274 | 0.288 | 0.360 | 0.345 | 0.318 | 0.299 | 0.289 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.316 |
| 2007 | 0.059 | 0.098 | 0.206 | 0.257 | 0.325 | 0.345 | 0.309 | 0.309 | 0.325 | 0.288 | 0.328 | 0.350 | 0.000 | 0.000 | 0.000 | 0.000 | 0.320 |
| 2008 | 0.027 | 0.104 | 0.218 | 0.282 | 0.315 | 0.402 | 0.407 | 0.317 | 0.359 | 0.337 | 0.334 | 0.433 | 0.000 | 0.000 | 0.000 | 0.000 | 0.354 |
| 2009 | 0.042 | 0.092 | 0.220 | 0.289 | 0.381 | 0.401 | 0.465 | 0.393 | 0.336 | 0.310 | 0.342 | 0.321 | 0.436 | 0.000 | 0.000 | 0.000 | 0.328 |
| 2010 | 0.022 | 0.088 | 0.226 | 0.305 | 0.376 | 0.448 | 0.422 | 0.458 | 0.380 | 0.376 | 0.351 | 0.355 | 0.272 | 0.000 | 0.000 | 0.000 | 0.373 |
| 2011 | 0.046 | 0.106 | 0.185 | 0.315 | 0.379 | 0.443 | 0.499 | 0.460 | 0.568 | 0.606 | 0.396 | 0.437 | 0.894 | 0.000 | 0.000 | 0.000 | 0.501 |
| 2012 | 0.021 | 0.086 | 0.191 | 0.275 | 0.376 | 0.391 | 0.403 | 0.413 | 0.437 | 0.583 | 0.223 | 0.473 | 0.616 | 0.489 | 0.288 | 0.000 | 0.458 |
| 2013 | 0.045 | 0.090 | 0.186 | 0.244 | 0.397 | 0.481 | 0.497 | 0.522 | 0.465 | 0.567 | 1.027 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.496 |
| 2014 | 0.023 | 0.111 | 0.212 | 0.289 | 0.328 | 0.472 | 0.499 | 0.527 | 0.606 | 0.623 | 0.611 | 1.754 | 0.000 | 0.000 | 0.000 | 0.000 | 0.612 |
| 2015 | 0.044 | 0.122 | 0.201 | 0.305 | 0.372 | 0.384 | 0.474 | 0.511 | 0.509 | 0.803 | 0.758 | 0.778 | 0.000 | 0.000 | 0.000 | 0.000 | 0.535 |

Table 12.9: Whiting in Subarea 4 and Division 7.d. Landings mean weights at age (kg). Age 8 is a plus-group. Ages 1-8+ and years 1990-2015 are included in the final assessment.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 0.000 | 0.185 | 0.233 | 0.250 | 0.334 | 0.426 | 0.434 | 0.466 | 0.615 | 0.702 | 1.539 | 0.589 | 0.000 | 0.000 | 0.000 | 0.000 | 0.649 |
| 1979 | 0.113 | 0.206 | 0.231 | 0.277 | 0.304 | 0.416 | 0.456 | 0.491 | 0.583 | 0.617 | 0.737 | 0.515 | 0.000 | 0.000 | 0.000 | 0.000 | 0.587 |
| 1980 | 0.000 | 0.204 | 0.239 | 0.273 | 0.335 | 0.358 | 0.47 | 0.45 | 0.568 | 0.539 | 0.790 | 0.688 | 1.711 | 0.000 | 0.000 | 0.000 | 0.572 |
| 1981 | 0.144 | 0.194 | 0.242 | 0.292 | 0.331 | 0.378 | 0.41 | 0.445 | 0.651 | 0.833 | 1.041 | 0.695 | 0.000 | 0.000 | 0.000 | 0.000 | 0.720 |
| 1982 | 0.000 | 0.186 | 0.230 | 0.282 | 0.340 | 0.396 | 0.46 | 0.507 | 0.702 | 0.772 | 1.14 | 0.853 | 1.081 | 0.000 | 0.000 | 0.000 | 0.736 |
| 1983 | 0.132 | 0.199 | 0.240 | 0.282 | 0.332 | 0.383 | 0.429 | 0.452 | 0.522 | 0.677 | 0.51 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.538 |
| 1984 | 0.000 | 0.194 | 0.231 | 0.279 | 0.346 | 0.391 | 0.403 | 0.472 | 0.575 | 0.514 | 0.871 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.567 |
| 1985 | 0.137 | 0.187 | 0.248 | 0.307 | 0.337 | 0.408 | 0.443 | 0.498 | 0.426 | 0.507 | 0.852 | 0.976 | 0.000 | 0.000 | 0.000 | 0.000 | 0.438 |
| 1986 | 0.131 | 0.189 | 0.230 | 0.279 | 0.327 | 0.376 | 0.484 | 0.472 | 0.546 | 1.226 | 0.990 | 0.535 | 0.000 | 0.000 | 0.000 | 0.000 | 0.632 |
| 1987 | 0.135 | 0.188 | 0.226 | 0.286 | 0.310 | 0.381 | 0.38 | 0.542 | 0.564 | 0.857 | 0.60 | 1.193 | 0.000 | 0.000 | 0.000 | 0.000 | 0.593 |
| 1988 | 0.117 | 0.194 | 0.22 | 0.256 | 0.328 | 0.351 | 0.42 | 0.50 | 0.88 | 0.585 | 0.64 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.702 |
| 19 | 0.17 | 0.17 | 0.2 | 0.25 | 0.288 | 0.345 | 0.37 | 0.44 | 0.373 | 0.5 | 0.85 | 0.6 | 0.000 | 0.000 | 0.000 | 0.000 | 0.405 |
| 199 | 0.167 | 0.206 | 0.2 | 0.26 | 0.2 | 0.33 | 0.45 | 0.53 | 0.6 | 0.35 | 0.9 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.597 |
| 19 | 0.13 | 0.202 | 0.24 | 0.25 | 0.30 | 0.3 | 0.34 | 0.38 | 0.385 | 0.58 | 0.99 | 2.756 | 0.000 | 0.000 | 0.000 | 0.000 | 0.401 |
| 1992 | 0.145 | 0.194 | 0.246 | 0.28 | 0.306 | 0.340 | 0.35 | 0.38 | 0.473 | 0.763 | 1.728 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.501 |
| 1993 | 0.153 | 0.194 | 0.248 | 0.284 | 0.345 | 0.358 | 0.38 | 0.418 | 0.376 | 0.417 | 0.359 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.379 |
| 1994 | 0.132 | 0.182 | 0.248 | 0.297 | 0.346 | 0.392 | 0.382 | 0.412 | 0.414 | 0.392 | 0.499 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.410 |
| 1995 | 0.140 | 0.171 | 0.256 | 0.299 | 0.367 | 0.397 | 0.43 | 0.437 | 0.448 | 0.346 | 0.406 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.421 |
| 1996 | 0.143 | 0.169 | 0.222 | 0.274 | 0.329 | 0.408 | 0.41 | 0.45 | 0.439 | 0.404 | 0.376 | 0.398 | 0.287 | 0.000 | 0.000 | 0.000 | 0.432 |
| 1997 | 0.149 | 0.171 | 0.206 | 0.260 | 0.315 | 0.349 | 0.40 | 0.386 | 0.398 | 0.479 | 0.43 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.424 |
| 199 | 0.138 | 0.164 | 0.208 | 0.25 | 0.304 | 0.331 | 0.36 | 0.34 | 0.392 | 0.504 | 0.60 | 0.600 | 0.000 | 0.000 | 0.000 | 0.000 | 0.427 |
| 199 | 0.135 | 0.184 | 0.23 | 0.27 | 0.281 | 0.303 | 0.31 | 0.32 | 0.292 | 0.368 | 0.344 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.301 |
| 2000 | 0.000 | 0.166 | 0.22 | 0.272 | 0.299 | 0.292 | 0.31 | 0.27 | 0.269 | 0.264 | 0.28 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.269 |
| 2001 | 0.138 | 0.160 | 0.216 | 0.268 | 0.285 | 0.267 | 0.30 | 0.288 | 0.287 | 0.301 | 0.315 | 0.505 | 0.000 | 0.000 | 0.000 | 0.000 | 0.293 |
| 2002 | 0.000 | 0.183 | 0.214 | 0.260 | 0.293 | 0.313 | 0.36 | 0.350 | 0.325 | 0.390 | 0.311 | 0.231 | 0.304 | 0.643 | 0.000 | 0.000 | 0.333 |
| 2003 | 0.128 | 0.208 | 0.228 | 0.258 | 0.308 | 0.311 | 0.374 | 0.391 | 0.342 | 0.462 | 0.620 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.369 |
| 2004 | 0.000 | 0.210 | 0.216 | 0.242 | 0.290 | 0.326 | 0.330 | 0.334 | 0.366 | 0.351 | 0.352 | 1.463 | 0.337 | 0.000 | 0.000 | 0.000 | 0.363 |
| 2005 | 0.164 | 0.205 | 0.253 | 0.277 | 0.270 | 0.308 | 0.339 | 0.313 | 0.296 | 0.381 | 0.316 | 0.337 | 0.670 | 0.000 | 0.000 | 0.000 | 0.313 |
| 2006 | 0.133 | 0.217 | 0.254 | 0.285 | 0.295 | 0.298 | 0.377 | 0.353 | 0.334 | 0.306 | 0.290 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.331 |
| 2007 | 0.202 | 0.199 | 0.264 | 0.280 | 0.351 | 0.361 | 0.319 | 0.332 | 0.342 | 0.318 | 0.334 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.338 |
| 2008 | 0.000 | 0.223 | 0.265 | 0.324 | 0.356 | 0.431 | 0.424 | 0.359 | 0.389 | 0.339 | 0.334 | 0.433 | 0.000 | 0.000 | 0.000 | 0.000 | 0.374 |
| 2009 | 0.148 | 0.205 | 0.246 | 0.318 | 0.386 | 0.404 | 0.464 | 0.404 | 0.347 | 0.313 | 0.311 | 0.321 | 0.436 | 0.000 | 0.000 | 0.000 | 0.329 |
| 2010 | 0.359 | 0.221 | 0.255 | 0.331 | 0.416 | 0.470 | 0.479 | 0.541 | 0.439 | 0.374 | 0.337 | 0.355 | 0.272 | 0.000 | 0.000 | 0.000 | 0.388 |
| 2011 | 0.000 | 0.182 | 0.237 | 0.374 | 0.416 | 0.506 | 0.569 | 0.504 | 0.582 | 0.634 | 0.406 | 0.465 | 0.894 | 0.000 | 0.000 | 0.000 | 0.523 |
| 2012 | 0.021 | 0.135 | 0.236 | 0.337 | 0.468 | 0.443 | 0.501 | 0.478 | 0.478 | 0.584 | 0.256 | 0.514 | 0.621 | 0.489 | 0.288 | 0.000 | 0.498 |
| 2013 | 0.066 | 0.181 | 0.224 | 0.275 | 0.421 | 0.487 | 0.508 | 0.526 | 0.464 | 0.567 | 1.027 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.496 |
| 2014 | 0.000 | 0.177 | 0.256 | 0.323 | 0.359 | 0.513 | 0.547 | 0.546 | 0.634 | 0.647 | 0.613 | 1.754 | 0.000 | 0.000 | 0.000 | 0.000 | 0.638 |
| 2015 | 0.047 | 0.166 | 0.240 | 0.359 | 0.414 | 0.437 | 0.510 | 0.528 | 0.579 | 0.834 | 0.758 | 0.778 | 0.000 | 0.000 | 0.000 | 0.000 | 0.604 |

Table 12.10: Whiting in Subarea 4 and Division 7.d. Discards mean weights at age (kg). Age 8 is a plus-group. Ages 1-8+ and years 1990-2015 are included in the final assessment.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 0.036 | 0.145 | 0.158 | 0.185 | 0.209 | 0.222 | 0.239 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1979 | 0.080 | 0.104 | 0.158 | 0.191 | 0.189 | 0.234 | 0.265 | 0.295 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1980 | 0.030 | 0.107 | 0.166 | 0.202 | 0.244 | 0.253 | 0.26 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1981 | 0.071 | 0.131 | 0.164 | 0.19 | 0.230 | 0.289 | 0.25 | 0.26 | 0.000 | 0.000 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1982 | 0.047 | 0.091 | 0.182 | 0.211 | 0.225 | 0.241 | 0.24 | 0.26 | 0.000 | 0.000 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1983 | 0.036 | 0.114 | 0.167 | 0.23 | 0.26 | 0.290 | 0.31 | 0.27 | 0.36 | 0.000 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.365 |
| 198 | 0.038 | 0.101 | 0.162 | 0.21 | 0.24 | 0.265 | 0.24 | 0.27 | 0.00 | 0.000 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1985 | 0.02 | 0.105 | 0.16 | 0.21 | 0.23 | 0.24 | 0.25 | 0.25 | 0.00 | 0.000 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1986 | 0.028 | 0.123 | 0.166 | 0.19 | 0.208 | 0.227 | 0.19 | 0.21 | 0.31 | 0.000 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.311 |
| 1987 | 0.016 | 0.090 | 0.149 | 0.206 | 0.205 | 0.263 | 0.25 | 0.000 | 0.292 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.292 |
| 1988 | 0.030 | 0.063 | 0.146 | 0.181 | 0.210 | 0.219 | 0.235 | 0.000 | 0.284 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.284 |
| 1989 | 0.033 | 0.083 | 0.164 | 0.19 | 0.213 | 0.227 | 0.24 | 0.35 | 0.221 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.221 |
| 1990 | 0.024 | 0.095 | 0.130 | 0.183 | 0.186 | 0.196 | 0.24 | 0.30 | 0.000 | 0.000 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 | 0.041 | 0.089 | 0.154 | 0.17 | 0.21 | 0.230 | 0.25 | 0.26 | 0.00 | 0.00 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 | 0.03 | 0.0 | 0.1 | 0.2 | 0.21 | 0.24 | 0.2 | 0. | 1.18 | 0.00 | 0.0 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 1.183 |
| 19 | 0.0 | 0.0 | 0.1 | 0.20 | 0. | 0.235 | 0.2 | 0. | 0. | 0.000 | 0.0 | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0 |
| 19 | 0.0 | 0.0 | 0.1 | 0.20 | 0.2 | 0.244 | 0.25 | 0.33 | 0.00 | 0.000 | 0.0 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 | 0.032 | 0.1 | 0.1 | 0.20 | 0. | 0.247 | 0.24 | 0.33 | 0.29 | 0.000 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.290 |
| 1996 | 0.031 | 0.094 | 0.151 | 0.19 | 0.225 | 0.281 | 0.26 | 0.30 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.031 | 0.125 | 0.181 | 0.21 | 0.225 | 0.233 | 0.25 | 0.61 | 0.320 | 0.601 | 0.773 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.352 |
| 1998 | 0.026 | 0.086 | 0.173 | 0.204 | 0.228 | 0.234 | 0.22 | 0.24 | 0.191 | 0.180 | 0.284 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.206 |
| 1999 | 0.062 | 0.100 | 0.166 | 0.19 | 0.20 | 0.225 | 0.23 | 0.21 | 0.231 | 0.220 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.227 |
| 2000 | 0.033 | 0.127 | 0.167 | 0.19 | 0.226 | 0.209 | 0.21 | 0.22 | 0.264 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.264 |
| 200 | 0.023 | 0.084 | 0.183 | 0.21 | 0.25 | 0.248 | 0.24 | 0.22 | 0.243 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.243 |
| 2002 | 0.039 | 0.130 | 0.167 | 0.19 | 0.22 | 0.224 | 0.22 | 0.27 | 0.334 | 1.120 | 0.217 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.352 |
| 2003 | 0.048 | 0.062 | 0.105 | 0.17 | 0.21 | 0.262 | 0.25 | 0.29 | 0.237 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.055 |
| 2004 | 0.079 | 0.131 | 0.158 | 0.203 | 0.223 | 0.239 | 0.23 | 0.22 | 0.204 | 0.351 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.245 |
| 2005 | 0.070 | 0.124 | 0.177 | 0.20 | 0.221 | 0.223 | 0.235 | 0.245 | 0.222 | 0.293 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.224 |
| 2006 | 0.093 | 0.131 | 0.161 | 0.193 | 0.229 | 0.233 | 0.247 | 0.273 | 0.239 | 0.279 | 0.289 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.246 |
| 2007 | 0.050 | 0.065 | 0.170 | 0.214 | 0.225 | 0.247 | 0.237 | 0.215 | 0.229 | 0.166 | 0.241 | 0.350 | 0.000 | 0.000 | 0.000 | 0.000 | 0.217 |
| 2008 | 0.027 | 0.072 | 0.181 | 0.213 | 0.230 | 0.265 | 0.32 | 0.244 | 0.291 | 0.317 | 0.057 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.293 |
| 2009 | 0.042 | 0.089 | 0.193 | 0.243 | 0.376 | 0.393 | 0.484 | 0.286 | 0.300 | 0.268 | 0.596 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.319 |
| 2010 | 0.019 | 0.075 | 0.211 | 0.272 | 0.319 | 0.384 | 0.330 | 0.254 | 0.290 | 0.390 | 0.730 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.323 |
| 2011 | 0.046 | 0.093 | 0.147 | 0.242 | 0.271 | 0.285 | 0.339 | 0.344 | 0.246 | 0.291 | 0.304 | 0.167 | 0.000 | 0.000 | 0.000 | 0.000 | 0.256 |
| 2012 | 0.021 | 0.076 | 0.155 | 0.184 | 0.204 | 0.254 | 0.211 | 0.252 | 0.313 | 0.574 | 0.183 | 0.374 | 0.580 | 0.489 | 0.288 | 0.000 | 0.325 |
| 2013 | 0.045 | 0.076 | 0.158 | 0.196 | 0.262 | 0.326 | 0.207 | 0.335 | 0.508 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.508 |
| 2014 | 0.023 | 0.103 | 0.170 | 0.237 | 0.263 | 0.295 | 0.340 | 0.418 | 0.393 | 0.343 | 0.509 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.381 |
| 2015 | 0.044 | 0.115 | 0.175 | 0.236 | 0.275 | 0.270 | 0.326 | 0.431 | 0.299 | 0.406 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.301 |

Table 12.11: Whiting in Subarea 4 and Division 7.d. Industrial bycatch mean weights at age (kg). Age 8 is a plus-group. Ages 1-8+ and years 1990-2015 are included in the final assessment.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 0.009 | 0.059 | 0.158 | 0.220 | 0.295 | 0.529 | 0.351 | 0.449 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1979 | 0.008 | 0.069 | 0.141 | 0.249 | 0.428 | 0.477 | 0.467 | 0.605 | 0.482 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.482 |
| 1980 | 0.013 | 0.051 | 0.164 | 0.281 | 0.412 | 0.380 | 0.38 | 0.561 | 0.000 | 1.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.000 |
| 1981 | 0.011 | 0.056 | 0.141 | 0.218 | 0.318 | 0.433 | 0.596 | 0.600 | 0.800 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.800 |
| 1982 | 0.025 | 0.038 | 0.133 | 0.232 | 0.320 | 0.366 | 0.67 | 0.284 | 0.800 | 1.00 | 1.200 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.840 |
| 19 | 0.012 | 0.058 | 0.148 | 0.311 | 0.431 | 0.651 | 0.56 | 0.60 | 0.800 | 1.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.802 |
| 19 | 0.018 | 0.053 | 0.17 | 0.2 | 0.343 | 0.390 | 0.22 | 0.60 | 0.800 | 1.0 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.896 |
| 1985 | 0.014 | 0.054 | 0.150 | 0.263 | 0.382 | 0.454 | 0.50 | 0.584 | 0.800 | 1.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.809 |
| 1986 | 0.014 | 0.054 | 0.15 | 0.262 | 0.381 | 0.455 | 0.50 | 0.600 | 0.800 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.800 |
| 1987 | 0.012 | 0.043 | 0.085 | 0.173 | 0.262 | 0.400 | 0.500 | 0.600 | 0.800 | 1.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.822 |
| 1988 | 0.012 | 0.050 | 0.115 | 0.197 | 0.245 | 0.380 | 0.500 | 0.600 | 0.800 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.800 |
| 1989 | 0.022 | 0.053 | 0.137 | 0.224 | 0.285 | 0.344 | 0.48 | 0.396 | 0.385 | 0.40 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.385 |
| 199 | 0.007 | 0.073 | 0.123 | 0.18 | 0.201 | 0.280 | 0.35 | 0.33 | 0.472 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.472 |
| 19 | 0.018 | 0.105 | 0.13 | 0.21 | 0.2 | 0.265 | 0.2 | 0. | 0.00 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 | 0.0 | 0.068 | 0.1 | 0.23 | 0.244 | 0.3 | 0. | 0.25 | 0.282 | 0.0 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.282 |
| 19 | 0.0 | 0.0 | 0.1 | 0.2 | 0.264 | 0.3 | 0.2 | 0.3 | 0.0 | 0.0 | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 | 0.0 | 0.055 | 0.13 | 0.25 | 0.388 | 0.52 | 0.55 | 0.44 | 0.555 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.555 |
| 19 | 0.00 | 0.072 | 0.1 | 0.3 | 0.373 | 0.51 | 0.00 | 0. | 0.000 | 0.00 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 0.016 | 0.064 | 0.15 | 0.23 | 0.233 | 0.347 | 0.250 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.012 | 0.051 | 0.145 | 0.252 | 0.321 | 0.348 | 0.588 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1998 | 0.015 | 0.049 | 0.115 | 0.220 | 0.304 | 0.286 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 0.013 | 0.027 | 0.07 | 0.144 | 0.194 | 0.286 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2000 | 0.038 | 0.051 | 0.166 | 0.242 | 0.289 | 0.339 | 0.00 | 0.588 | 0.000 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 200 | 0.012 | 0.055 | 0.11 | 0.225 | 0.320 | 0.351 | 0.38 | 0.000 | 0.000 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2002 | 0.010 | 0.044 | 0.10 | 0.185 | 0.294 | 0.415 | 0.38 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2003 | 0.010 | 0.035 | 0.102 | 0.18 | 0.302 | 0.418 | 0.46 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2004 | 0.010 | 0.032 | 0.083 | 0.143 | 0.264 | 0.000 | 0.38 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.014 | 0.043 | 0.133 | 0.196 | 0.205 | 0.366 | 0.43 | 0.541 | 0.530 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.530 |
| 2006 | 0.000 | 0.046 | 0.119 | 0.208 | 0.277 | 0.362 | 0.40 | 0.564 | 0.530 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.530 |
| 2007 | 0.000 | 0.046 | 0.119 | 0.208 | 0.277 | 0.362 | 0.401 | 0.564 | 0.530 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.530 |
| 2008 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 0.000 | 0.046 | 0.119 | 0.208 | 0.277 | 0.362 | 0.401 | 0.564 | 0.530 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.530 |
| 2010 | 0.000 | 0.046 | 0.119 | 0.208 | 0.277 | 0.362 | 0.401 | 0.564 | 0.530 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.530 |
| 2011 | 0.000 | 0.188 | 0.242 | 0.305 | 0.321 | 0.371 | 0.464 | 0.436 | 0.628 | 0.421 | 0.393 | 0.480 | 0.000 | 0.000 | 0.000 | 0.000 | 0.419 |
| 2012 | 0.021 | 0.087 | 0.193 | 0.279 | 0.385 | 0.398 | 0.413 | 0.421 | 0.443 | 0.583 | 0.225 | 0.478 | 0.618 | 0.489 | 0.288 | 0.000 | 0.462 |
| 2013 | 0.045 | 0.090 | 0.186 | 0.244 | 0.397 | 0.481 | 0.497 | 0.522 | 0.465 | 0.567 | 1.027 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.497 |
| 2014 | 0.023 | 0.111 | 0.212 | 0.289 | 0.328 | 0.472 | 0.499 | 0.527 | 0.606 | 0.623 | 0.611 | 1.754 | 0.000 | 0.000 | 0.000 | 0.000 | 0.613 |
| 2015 | 0.044 | 0.122 | 0.201 | 0.305 | 0.372 | 0.384 | 0.474 | 0.511 | 0.509 | 0.803 | 0.758 | 0.778 | 0.000 | 0.000 | 0.000 | 0.000 | 0.538 |

Table 12.12: Whiting in Subarea 4 and Division 7.d. Estimated proportion mature at age as used in the assessment.

| Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mat | 0.11 | 0.92 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 12.13: Whiting in Subarea 4 and Division 7.d. Natural mortality at age from ICES-WGSAM (2014).

| Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 1.248 | 0.608 | 0.549 | 0.545 | 0.478 | 0.478 | 0.478 | 0.478 |
| 1991 | 1.242 | 0.604 | 0.547 | 0.544 | 0.487 | 0.487 | 0.487 | 0.487 |
| 1992 | 1.237 | 0.603 | 0.546 | 0.543 | 0.496 | 0.496 | 0.496 | 0.496 |
| 1993 | 1.233 | 0.603 | 0.546 | 0.543 | 0.506 | 0.506 | 0.506 | 0.506 |
| 1994 | 1.233 | 0.605 | 0.547 | 0.544 | 0.515 | 0.515 | 0.515 | 0.515 |
| 1995 | 1.238 | 0.607 | 0.549 | 0.545 | 0.523 | 0.523 | 0.523 | 0.523 |
| 1996 | 1.246 | 0.609 | 0.551 | 0.547 | 0.531 | 0.531 | 0.531 | 0.531 |
| 1997 | 1.258 | 0.610 | 0.553 | 0.549 | 0.538 | 0.538 | 0.538 | 0.538 |
| 1998 | 1.274 | 0.612 | 0.556 | 0.551 | 0.544 | 0.544 | 0.544 | 0.544 |
| 1999 | 1.292 | 0.614 | 0.558 | 0.552 | 0.549 | 0.549 | 0.549 | 0.549 |
| 2000 | 1.314 | 0.619 | 0.562 | 0.555 | 0.554 | 0.554 | 0.554 | 0.554 |
| 2001 | 1.338 | 0.626 | 0.567 | 0.559 | 0.559 | 0.559 | 0.559 | 0.559 |
| 2002 | 1.362 | 0.637 | 0.574 | 0.565 | 0.566 | 0.566 | 0.566 | 0.566 |
| 2003 | 1.380 | 0.651 | 0.583 | 0.573 | 0.573 | 0.573 | 0.573 | 0.573 |
| 2004 | 1.386 | 0.668 | 0.592 | 0.582 | 0.579 | 0.579 | 0.579 | 0.579 |
| 2005 | 1.379 | 0.686 | 0.601 | 0.591 | 0.584 | 0.584 | 0.584 | 0.584 |
| 2006 | 1.362 | 0.704 | 0.608 | 0.599 | 0.586 | 0.586 | 0.586 | 0.586 |
| 2007 | 1.338 | 0.722 | 0.613 | 0.605 | 0.585 | 0.585 | 0.585 | 0.585 |
| 2008 | 1.312 | 0.739 | 0.617 | 0.610 | 0.580 | 0.580 | 0.580 | 0.580 |
| 2009 | 1.288 | 0.755 | 0.620 | 0.615 | 0.574 | 0.574 | 0.574 | 0.574 |
| 2010 | 1.271 | 0.771 | 0.624 | 0.620 | 0.567 | 0.567 | 0.567 | 0.567 |
| 2011 | 1.261 | 0.790 | 0.629 | 0.627 | 0.561 | 0.561 | 0.561 | 0.561 |
| 2012 | 1.257 | 0.809 | 0.636 | 0.635 | 0.557 | 0.557 | 0.557 | 0.557 |
| 2013 | 1.255 | 0.830 | 0.643 | 0.644 | 0.553 | 0.553 | 0.553 | 0.553 |
| 2014 | 1.255 | 0.830 | 0.643 | 0.644 | 0.553 | 0.553 | 0.553 | 0.553 |
| 2015 | 1.255 | 0.830 | 0.643 | 0.644 | 0.553 | 0.553 | 0.553 | 0.553 |

Table 12.14: Whiting in Subarea 4 and Division 7.d. Tuning series used in the assessment and forecast. Note that only years from 1990 onwards are used in the final assessment.

| IBTS-Q1 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 1 | 2 | 3 | 4 | 5 |
| 1978 | 5.472 | 2.629 | 0.919 | 0.220 | 0.042 |
| 1979 | 4.439 | 2.307 | 1.143 | 0.335 | 0.050 |
| 1980 | 6.750 | 4.037 | 1.250 | 0.254 | 0.088 |
| 1981 | 2.297 | 4.635 | 2.285 | 0.460 | 0.091 |
| 1982 | 1.515 | 2.173 | 2.581 | 0.686 | 0.101 |
| 1983 | 1.266 | 1.250 | 1.100 | 0.764 | 0.322 |
| 1984 | 4.345 | 1.780 | 0.890 | 0.303 | 0.254 |
| 1985 | 3.392 | 3.623 | 0.659 | 0.186 | 0.071 |
| 1986 | 4.687 | 2.683 | 1.946 | 0.321 | 0.066 |
| 1987 | 6.849 | 5.611 | 0.904 | 0.455 | 0.049 |
| 1988 | 4.480 | 8.657 | 3.143 | 0.330 | 0.126 |
| 1989 | 14.476 | 5.328 | 4.055 | 1.073 | 0.119 |
| 1990 | 5.189 | 8.624 | 1.982 | 0.916 | 0.169 |
| 1991 | 10.076 | 6.864 | 4.796 | 0.709 | 0.376 |
| 1992 | 9.073 | 6.657 | 2.402 | 1.508 | 0.127 |
| 1993 | 10.756 | 5.228 | 2.446 | 0.655 | 0.590 |
| 1994 | 7.217 | 6.274 | 1.810 | 0.681 | 0.119 |
| 1995 | 6.786 | 4.485 | 2.394 | 0.581 | 0.119 |
| 1996 | 5.024 | 4.860 | 2.447 | 0.697 | 0.231 |
| 1997 | 2.878 | 3.422 | 1.624 | 0.604 | 0.180 |
| 1998 | 5.431 | 1.607 | 1.254 | 0.540 | 0.155 |
| 1999 | 6.763 | 3.054 | 0.947 | 0.575 | 0.258 |
| 2000 | 7.658 | 5.444 | 1.835 | 0.536 | 0.202 |
| 2001 | 6.487 | 5.984 | 2.991 | 0.983 | 0.258 |
| 2002 | 5.574 | 3.433 | 2.632 | 0.633 | 0.208 |
| 2003 | 1.316 | 2.988 | 2.370 | 1.334 | 0.484 |
| 2004 | 1.844 | 0.901 | 1.727 | 0.999 | 0.487 |
| 2005 | 1.127 | 0.978 | 0.456 | 0.601 | 0.390 |
| 2006 | 1.844 | 1.251 | 0.455 | 0.183 | 0.270 |
| 2007 | 0.645 | 1.473 | 0.673 | 0.186 | 0.084 |
| 2008 | 2.686 | 2.058 | 0.655 | 0.221 | 0.075 |
| 2009 | 2.112 | 2.958 | 0.936 | 0.272 | 0.119 |
| 2010 | 3.262 | 2.248 | 2.441 | 0.948 | 0.285 |
| 2011 | 1.849 | 3.371 | 1.575 | 0.926 | 0.197 |
| 2012 | 2.313 | 5.883 | 1.147 | 0.464 | 0.324 |
| 2013 | 0.544 | 1.630 | 2.413 | 0.883 | 0.269 |
| 2014 | 2.652 | 1.845 | 0.992 | 0.659 | 0.227 |
| 2015 | 3.150 | 2.126 | 0.598 | 0.287 | 0.240 |


| 2016 | 2.813 | 3.122 | $0.884 \quad 0.197$ | 0.111 |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| IBTS-Q3 |  |  |  |  |  |
| Age | 1 | 2 | 3 | 4 | 5 |
| 1991 | 7.034 | 1.586 | 0.790 | 0.146 | 0.052 |
| 1992 | 6.009 | 2.961 | 0.725 | 0.575 | 0.103 |
| 1993 | 6.387 | 1.774 | 0.661 | 0.147 | 0.159 |
| 1994 | 6.776 | 2.195 | 0.747 | 0.195 | 0.047 |
| 1995 | 6.198 | 2.912 | 1.072 | 0.215 | 0.060 |
| 1996 | 5.457 | 2.782 | 1.294 | 0.340 | 0.069 |
| 1997 | 3.330 | 1.807 | 1.090 | 0.280 | 0.107 |
| 1998 | 3.306 | 1.502 | 0.528 | 0.310 | 0.112 |
| 1999 | 12.035 | 1.906 | 0.539 | 0.245 | 0.095 |
| 2000 | 9.417 | 3.269 | 0.641 | 0.136 | 0.065 |
| 2001 | 6.450 | 2.823 | 0.949 | 0.193 | 0.043 |
| 2002 | 7.321 | 2.374 | 1.251 | 0.340 | 0.053 |
| 2003 | 2.462 | 3.021 | 1.348 | 0.661 | 0.165 |
| 2004 | 1.506 | 0.590 | 0.663 | 0.457 | 0.271 |
| 2005 | 1.714 | 0.683 | 0.314 | 0.456 | 0.340 |
| 2006 | 1.746 | 0.863 | 0.326 | 0.135 | 0.233 |
| 2007 | 0.955 | 0.636 | 0.376 | 0.115 | 0.084 |
| 2008 | 3.623 | 0.689 | 0.309 | 0.138 | 0.041 |
| 2009 | 5.855 | 3.848 | 0.410 | 0.123 | 0.080 |
| 2010 | 2.243 | 1.457 | 0.546 | 0.128 | 0.060 |
| 2011 | 4.468 | 1.444 | 0.472 | 0.162 | 0.069 |
| 2012 | 2.567 | 1.935 | 0.570 | 0.201 | 0.106 |
| 2013 | 0.675 | 0.601 | 0.658 | 0.175 | 0.071 |
| 2014 | 2.234 | 0.980 | 0.656 | 0.333 | 0.103 |
| 2015 | 3.125 | 2.226 | 0.431 | 0.240 | 0.184 |

Table 12.15: Whiting in Subarea 4 and Division 7.d. FLXSA tuning diagnostics.

FLR XSA Diagnostics 2016-04-27 10:43:51
CPUE data from indices
Catch data for 26 years 1990 to 2015. Ages 1 to 8 .
fleet first age last age first year last year alpha beta
1 IBTS-Q1 $1 \begin{array}{lllllll}1990 & 2015 & 0 & 0.25\end{array}$
2 IBTS-Q3 $1 \begin{array}{lllllll} & 1 & 5 & 1991 & 2015 & 0.5 & 0.75\end{array}$

Time series weights :
Tapered time weighting not applied

Catchability analysis :
Catchability independent of size for all ages
Catchability independent of age for ages $>4$

Terminal population estimation :
Survivor estimates shrunk towards the mean F of the final 3 years or the 4 oldest ages.
S.E. of the mean to which the estimates are shrunk $=2$

Minimum standard error for population estimates derived from each fleet $=0.3$ prior weighting not applied

Regression weights
year
age 2006200720082009201020112012201320142015
$\begin{array}{lllllllllll}\text { all } & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$

Fishing mortalities
year
age 2006200720082009201020112012201320142015
10.0960 .0510 .0440 .0170 .0210 .0200 .0340 .0330 .0390 .036
20.1720 .1400 .1140 .1220 .1560 .1480 .0530 .0610 .1050 .169
30.2870 .2420 .2190 .1340 .1210 .1690 .1290 .1240 .1740 .170

```
40.2350.3270.321 0.263 0.155 0.156 0.251 0.210 0.201 0.171
5 0.272 0.251 0.3620.3130.2910.134 0.144 0.397 0.339 0.284
6 0 . 3 8 9 0 . 3 3 6 0 . 2 6 8 ~ 0 . 2 9 4 ~ 0 . 4 3 1 0 . 2 7 5 ~ 0 . 2 0 0 ~ 0 . 1 6 1 ~ 0 . 3 7 9 ~ 0 . 3 4 3 )
70.3110.4780.461 0.200 0.251 0.417 0.369 0.1340.113 0.320
80.3110.4780.461 0.200 0.251 0.417 0.369 0.134 0.113 0.320
XSA population number (Thousand)
age
\(\begin{array}{lllllllll}\text { year } & 1 & 2 & 3 & 4 & 5 & 6 & 7 & 8\end{array}\)
20061600655378759117641557738475449206231397091
200715263693727251576444807724232359191856410585
200827672403805721573996705018927105111430110298 2009226346871297216227268204264137372450120362 2010233577761354729659376312283361087630958802 20113053030641257242598140727351581200740081175 20121612011847298251084109229643381753852032269 20131140360443461357534116920450213191082275185 201421556723145231818261660564975517412156214562 2015209775759135212350480344713452039368522291
```

Estimated population abundance at 1st Jan 2016
age
$\begin{array}{lllllllll}\text { year } & 1 & 2 & 3 & 4 & 5 & 6 & 7 & 8\end{array}$
2016057687921767654763355593089983242862

Fleet: IBTS-Q1

Log catchability residuals.
year
age 199019911992199319941995199619971998199920002001200220032004 $1 \quad 0.0550 .7210 .6600 .6960 .3610 .4070 .4460 .1590 .4710 .2090 .1550 .1710 .175-0.210$ 0.108
$2-0.1320 .5160 .4110 .2660 .2830 .0010 .1780 .159-0.3240 .0230 .1500 .022-0.336-0.280-$ 0.306
$3 \quad 0.0540 .1320 .2930 .1830 .0050 .050$ 0.115-0.220-0.170-0.197 0.239 0.229-0.195-0.104 -0.119

```
4-0.005 0.285 0.008 0.085 0.116-0.013-0.102 -0.240-0.312 -0.028 0.241 0.643-0.466 0.022
-0.103
5 -0.484 0.416 -0.383 0.091-0.448 -0.407 0.068-0.433-0.642 -0.129 -0.075 0.491-0.067
0.070-0.182
year
age 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015
1-0.655-0.161-1.171-0.344-0.389 0.012 -0.825 0.039-1.062 -0.114 0.085
2-0.288-0.336-0.158 0.155-0.107 -0.225 0.137 0.406-0.227 0.246-0.236
3-0.321-0.301-0.206-0.233 0.083 0.437 0.206-0.150 0.241 0.034-0.086
4-0.337-0.356-0.183-0.343-0.157 0.967 0.333-0.092 0.478-0.167-0.274
5-0.270-0.383-0.306-0.153-0.032 0.765 0.161 0.055 0.256-0.019-0.330
```

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

```
    1}22\mp@code{3
Mean_Logq -13.3366 -12.1779 -12.0530-12.1675-12.1675
S.E_Logq 0.3385}00.33850.3385 0.3385 0.3385
```

Fleet: IBTS-Q3
Log catchability residuals.

## year

age 199119921993199419951996199719981999200020012002200320042005 $10.2450 .1590 .0730 .1910 .2050 .4130 .193-0.1260 .7040 .2620 .0740 .3690 .406-0.135-$ 0.289

2 -0.307 $0.203-0.193-0.1810 .1520 .1930 .0870 .1580 .1370 .212-0.202-0.1770 .322-0.165$ -0.044
$3-0.7880 .001-0.1550 .0520 .1560 .3700 .256-0.2140 .1100 .078-0.122-0.118$ 0.135-0.286 0.113
$4-0.2740 .035-0.342-0.029-0.0150 .181-0.0410 .0770 .099-0.149-0.072-0.1940 .210-$ 0.0080 .247
$5-0.4000 .378-0.142-0.1950 .005-0.1370 .058$ 0.018-0.125-0.140 -0.307-0.537-0.121 0.1460 .477
year
age 2006200720082009201020112012201320142015
1 -0.261-0.859-0.141 0.509-0.490-0.075 0.015-0.977-0.413-0.052
$2-0.102-0.401-0.3480 .760-0.030-0.076-0.109-0.6130 .2460 .476$
$3 \quad 0.2440 .070-0.1370 .064-0.257-0.169-0.037-0.2450 .4580 .423$

```
\(4 \quad 0.2470 .2940 .143-0.021-0.154-0.528\) 0.004-0.221 0.064 0.447
50.3880 .6070 .1980 .508 0.127-0.045-0.217-0.113 0.1270 .313
```

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

```
    1 2 3 4 5
Mean_Logq -12.5625 -12.3436-12.4826 -12.6584 -12.6584
S.E_Logq 0.2959 0.2959 0.2959 0.2959 0.2959
```

Terminal year survivor and F summaries:
,Age 1 Year class =2014
source
scaledWts survivors yrcls
IBTS-Q1 $0.378 \quad 6278902014$
IBTS-Q3 0.5965476042014
fshk 0.0265554392014
,Age 2 Year class =2013
source
scaledWts survivors yrcls
IBTS-Q1 0.4981719902013
IBTS-Q3 0.4893501952013
fshk 0.0135117052013
,Age 3 Year class =2012
source
scaledWts survivors yrcls
IBTS-Q1 $0.493 \quad 502542012$
IBTS-Q3 0.493835842012
fshk $0.013 \quad 64962 \quad 2012$
,Age 4 Year class =2011
source
scaledWts survivors yrcls
$\begin{array}{llll}\text { IBTS-Q1 } & 0.432 \quad 27044 \quad 2011\end{array}$
IBTS-Q3 0.55355575 ..... 2011
fshk 0.01526094 ..... 2011
,Age 5 Year class $=2010$
source
scaledWts survivors yrcls
IBTS-Q1 0.415222052010
IBTS-Q3 0.568422672010
fshk 0.017289642010
,Age 6 Year class $=2009$
source
scaledWts survivors yrcls
fshk 1118722009
,Age 7 Year class =2008
source
scaledWts survivors yrcls
fshk 138512008

Table 12.16: Whiting in Subarea 4 and Division 7.d. Final fishing mortality estimates from FLXSA.

| Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 0.144 | 0.378 | 0.605 | 0.760 | 0.983 | 0.710 | 0.786 | 0.786 |
| 1991 | 0.074 | 0.349 | 0.359 | 0.523 | 0.865 | 0.625 | 0.608 | 0.608 |
| 1992 | 0.133 | 0.270 | 0.407 | 0.458 | 0.464 | 0.876 | 1.131 | 1.131 |
| 1993 | 0.112 | 0.310 | 0.534 | 0.607 | 0.668 | 0.394 | 0.740 | 0.740 |
| 1994 | 0.102 | 0.236 | 0.457 | 0.683 | 0.851 | 0.984 | 0.301 | 0.301 |
| 1995 | 0.087 | 0.226 | 0.410 | 0.455 | 0.679 | 0.963 | 1.215 | 1.215 |
| 1996 | 0.071 | 0.204 | 0.375 | 0.473 | 0.495 | 0.707 | 1.160 | 1.160 |
| 1997 | 0.066 | 0.192 | 0.338 | 0.404 | 0.502 | 0.299 | 0.482 | 0.482 |
| 1998 | 0.073 | 0.155 | 0.228 | 0.358 | 0.449 | 0.491 | 0.294 | 0.294 |
| 1999 | 0.094 | 0.225 | 0.322 | 0.429 | 0.471 | 0.532 | 0.429 | 0.429 |
| 2000 | 0.034 | 0.193 | 0.362 | 0.424 | 0.588 | 0.693 | 1.226 | 1.226 |
| 2001 | 0.026 | 0.096 | 0.168 | 0.287 | 0.440 | 0.538 | 0.589 | 0.589 |
| 2002 | 0.030 | 0.088 | 0.206 | 0.241 | 0.257 | 0.242 | 0.475 | 0.475 |
| 2003 | 0.148 | 0.199 | 0.164 | 0.225 | 0.221 | 0.164 | 0.119 | 0.119 |
| 2004 | 0.082 | 0.131 | 0.128 | 0.191 | 0.266 | 0.298 | 0.200 | 0.200 |
| 2005 | 0.061 | 0.190 | 0.150 | 0.148 | 0.204 | 0.254 | 0.257 | 0.257 |
| 2006 | 0.096 | 0.172 | 0.287 | 0.235 | 0.272 | 0.389 | 0.311 | 0.311 |
| 2007 | 0.051 | 0.140 | 0.242 | 0.327 | 0.251 | 0.336 | 0.478 | 0.478 |
| 2008 | 0.044 | 0.114 | 0.219 | 0.321 | 0.362 | 0.268 | 0.461 | 0.461 |
| 2009 | 0.017 | 0.122 | 0.134 | 0.263 | 0.313 | 0.294 | 0.200 | 0.200 |
| 2010 | 0.021 | 0.156 | 0.121 | 0.155 | 0.291 | 0.431 | 0.251 | 0.251 |
| 2011 | 0.020 | 0.148 | 0.169 | 0.156 | 0.134 | 0.275 | 0.417 | 0.417 |
| 2012 | 0.034 | 0.053 | 0.129 | 0.251 | 0.144 | 0.200 | 0.369 | 0.369 |
| 2013 | 0.033 | 0.061 | 0.124 | 0.210 | 0.397 | 0.161 | 0.134 | 0.134 |
| 2014 | 0.039 | 0.105 | 0.174 | 0.201 | 0.339 | 0.379 | 0.113 | 0.113 |
| 2015 | 0.036 | 0.169 | 0.170 | 0.171 | 0.284 | 0.343 | 0.320 | 0.320 |

Table 12.17: Whiting in Subarea 4 and Division 7.d. Final abundance estimates from FLXSA.

| Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 3602988 | 2158961 | 370932 | 207681 | 63125 | 4829 | 1677 | 245 |
| 1991 | 3561689 | 895596 | 805842 | 116943 | 56315 | 14652 | 1472 | 724 |
| 1992 | 3429482 | 955318 | 345389 | 325450 | 40238 | 14574 | 4818 | 201 |
| 1993 | 3911169 | 871714 | 398905 | 133160 | 119663 | 15403 | 3696 | 1781 |
| 1994 | 3665192 | 1019106 | 349554 | 135406 | 42163 | 36986 | 6262 | 2499 |
| 1995 | 3286257 | 964585 | 439502 | 128009 | 39711 | 10754 | 8262 | 815 |
| 1996 | 2338049 | 873564 | 419423 | 168388 | 47053 | 11937 | 2434 | 1957 |
| 1997 | 1785364 | 626229 | 387550 | 166060 | 60680 | 16865 | 3462 | 1508 |
| 1998 | 2473817 | 474935 | 280846 | 158888 | 64011 | 21448 | 7301 | 2037 |
| 1999 | 4023722 | 643315 | 220454 | 128218 | 64059 | 23727 | 7617 | 2453 |
| 2000 | 4784856 | 1006166 | 277821 | 91393 | 48072 | 23107 | 8045 | 2207 |
| 2001 | 3996189 | 1243444 | 446909 | 110298 | 34365 | 15347 | 6641 | 3623 |
| 2002 | 3432777 | 1021342 | 604257 | 214279 | 47360 | 12646 | 5125 | 2557 |
| 2003 | 1210399 | 853244 | 494886 | 276781 | 95664 | 20796 | 5639 | 4640 |
| 2004 | 1225220 | 262589 | 364723 | 234492 | 124607 | 43274 | 9957 | 2208 |
| 2005 | 1599745 | 282356 | 118192 | 177496 | 108247 | 53525 | 18000 | 3768 |
| 2006 | 1600655 | 378759 | 117641 | 55773 | 84754 | 49206 | 23139 | 7091 |
| 2007 | 1526369 | 372725 | 157644 | 48077 | 24232 | 35919 | 18564 | 10585 |
| 2008 | 2767240 | 380572 | 157399 | 67050 | 18927 | 10511 | 14301 | 10298 |
| 2009 | 2263468 | 712972 | 162272 | 68204 | 26413 | 7372 | 4501 | 20362 |
| 2010 | 2335777 | 613547 | 296593 | 76312 | 28336 | 10876 | 3095 | 8802 |
| 2011 | 3053030 | 641257 | 242598 | 140727 | 35158 | 12007 | 4008 | 1175 |
| 2012 | 1612011 | 847298 | 251084 | 109229 | 64338 | 17538 | 5203 | 2269 |
| 2013 | 1140360 | 443461 | 357534 | 116920 | 45021 | 31910 | 8227 | 5185 |
| 2014 | 2155672 | 314523 | 181826 | 166056 | 49755 | 17412 | 15621 | 4562 |
| 2015 | 2097757 | 591352 | 123504 | 80344 | 71345 | 20393 | 6852 | 2291 |

Table 12.18: Whiting in Subarea 4 and Division 7.d. Final FLXSA summary table. Units are millions of individuals and tonnes.

| Year | Recruitment | TSB | SSB | Catch | Landings | Discards | Bycatch | Yield /SSB | Mean F(26) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 3602988 | 747248 | 455397 | 152602 | 45662 | 55603 | 51337 | 0.100 | 0.687 |
| 1991 | 3561689 | 751778 | 411205 | 126742 | 51929 | 35058 | 39755 | 0.126 | 0.544 |
| 1992 | 3429482 | 665749 | 393038 | 108555 | 50946 | 32564 | 25045 | 0.130 | 0.495 |
| 1993 | 3911169 | 625893 | 359374 | 116911 | 51818 | 44370 | 20723 | 0.144 | 0.503 |
| 1994 | 3665192 | 645577 | 357759 | 101650 | 48486 | 35692 | 17473 | 0.136 | 0.642 |
| 1995 | 3286257 | 646234 | 372599 | 105494 | 45938 | 32176 | 27379 | 0.123 | 0.547 |
| 1996 | 2338049 | 540570 | 332860 | 76123 | 40503 | 30505 | 5116 | 0.122 | 0.451 |
| 1997 | 1785364 | 453710 | 292840 | 61435 | 35563 | 19660 | 6213 | 0.121 | 0.347 |
| 1998 | 2473817 | 450586 | 244557 | 47475 | 28288 | 15693 | 3494 | 0.116 | 0.336 |
| 1999 | 4023722 | 538961 | 250537 | 60845 | 30130 | 25677 | 5038 | 0.120 | 0.396 |
| 2000 | 4784856 | 856619 | 344713 | 63806 | 28583 | 26063 | 9160 | 0.083 | 0.452 |
| 2001 | 3996189 | 801579 | 421586 | 45242 | 25061 | 19237 | 944 | 0.059 | 0.306 |
| 2002 | 3432777 | 606256 | 383340 | 46450 | 20675 | 18501 | 7275 | 0.054 | 0.207 |
| 2003 | 1210399 | 376566 | 307048 | 45640 | 16161 | 26745 | 2734 | 0.053 | 0.194 |
| 2004 | 1225220 | 364002 | 239720 | 33557 | 13295 | 19048 | 1214 | 0.055 | 0.202 |
| 2005 | 1599745 | 380442 | 199482 | 28883 | 15471 | 12525 | 888 | 0.078 | 0.189 |
| 2006 | 1600655 | 373156 | 180601 | 37038 | 18535 | 16310 | 2193 | 0.103 | 0.271 |
| 2007 | 1526369 | 310686 | 171915 | 27125 | 18915 | 6971 | 1239 | 0.110 | 0.259 |
| 2008 | 2767240 | 456552 | 193429 | 28247 | 17951 | 10296 | 0 | 0.093 | 0.257 |
| 2009 | 2263468 | 461318 | 262550 | 27139 | 18418 | 7705 | 1016 | 0.070 | 0.225 |
| 2010 | 2335777 | 484244 | 291072 | 31147 | 18224 | 11577 | 1346 | 0.063 | 0.231 |
| 2011 | 3053030 | 595720 | 298456 | 32626 | 18899 | 11977 | 1750 | 0.063 | 0.176 |
| 2012 | 1612011 | 446216 | 309860 | 25078 | 17032 | 7968 | 78 | 0.055 | 0.155 |
| 2013 | 1140360 | 363584 | 265474 | 26841 | 19335 | 5976 | 1530 | 0.073 | 0.191 |
| 2014 | 2155672 | 455526 | 237652 | 30675 | 18746 | 10451 | 1479 | 0.079 | 0.239 |
| 2015 | 2097757 | 484595 | 246870 | 33188 | 17707 | 13428 | 2053 | 0.072 | 0.227 |

Table 12.19: Whiting in Subarea 4 and Division 7.d. Reference points as determined in the Interbenchmark 2016 (ICES 2016).

| Reference point | value |
| :--- | :--- |
| Blim | 172741 (Bloss) |
| Flim | 0.39 |
| Bpa | 241837 (Btrigger) |
| Fpa | 0.28 |
| Fp. 05 (without Btrigger) | 0.12 |
| Fp. 05 (with Btrigger) | 0.15 (final Fmsy , with SSB $>$ Blim) |

Table 12.20: Whiting in Subarea 4 and Division 7.d. RCT3 input table.

| Year <br> class | Recruitment | IBTSQ1 1 | IBTSQ1 2 | IBTSQ30 | IBTSQ3 1 | IBTSQ32 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 3602988 | 518.936 | 686.445 | NA | NA | 158.594 |
| 1990 | 3561689 | 1007.621 | 665.714 | NA | 703.368 | 296.100 |
| 1991 | 3429482 | 907.297 | 522.811 | 536.990 | 600.867 | 177.377 |
| 1992 | 3911169 | 1075.624 | 627.406 | 1379.459 | 638.722 | 219.541 |
| 1993 | 3665192 | 721.709 | 448.484 | 919.193 | 677.645 | 291.180 |
| 1994 | 3286257 | 678.590 | 485.968 | 610.743 | 619.786 | 278.218 |
| 1995 | 2338049 | 502.361 | 342.246 | 729.246 | 545.708 | 180.681 |
| 1996 | 1785364 | 287.779 | 160.695 | 316.501 | 332.968 | 150.205 |
| 1997 | 2473817 | 543.117 | 305.445 | 2062.670 | 330.600 | 190.643 |
| 1998 | 4023722 | 676.266 | 544.367 | 2631.690 | 1203.501 | 326.943 |
| 1999 | 4784856 | 765.844 | 598.356 | 2498.550 | 941.658 | 282.320 |
| 2000 | 3996189 | 648.657 | 343.308 | 1968.070 | 645.003 | 237.372 |
| 2001 | 3432777 | 557.353 | 298.849 | 3031.442 | 732.137 | 302.054 |
| 2002 | 1210399 | 131.599 | 90.134 | 264.063 | 246.155 | 59.032 |
| 2003 | 1225220 | 184.399 | 97.824 | 363.406 | 150.623 | 68.259 |
| 2004 | 1599745 | 112.663 | 125.057 | 714.270 | 171.386 | 86.336 |
| 2005 | 1600655 | 184.411 | 147.304 | 169.321 | 174.625 | 63.592 |
| 2006 | 1526369 | 64.530 | 205.798 | 198.949 | 95.495 | 68.886 |
| 2007 | 2767240 | 268.598 | 295.812 | 822.902 | 362.299 | 384.777 |
| 2008 | 2263468 | 211.202 | 224.795 | 764.759 | 585.529 | 145.671 |
| 2009 | 2335777 | 326.192 | 337.096 | 593.801 | 224.321 | 144.439 |
| 2010 | 3053030 | 184.867 | 588.309 | 510.123 | 446.812 | 193.523 |
| 2011 | 1612011 | 231.255 | 162.985 | 247.085 | 256.718 | 60.102 |
| 2012 | 1140360 | 54.431 | 184.517 | 306.812 | 67.451 | 97.962 |
| 2013 | NA | 265.226 | 212.642 | 334.257 | 223.400 | 222.551 |
| 2014 | NA | 315.019 | 312.194 | 1401.008 | 312.453 | NA |
| 2015 | NA | 281.272 | NA | 2091.636 | NA | NA |

Table 12.21: Whiting in Subarea 4 and Division 7.d. RCT3 output.

| ANALYSIS BY RCT3 vER4.0 |  |  |  |
| :--- | :--- | :--- | :--- |
| Whiting |  |  |  |

Table 12.22: Whiting in Subarea 4 and Division 7.d. Recruitment estimates as used in the short-term forecast (RCT3 estimate year class 2015, geometric mean).

| Year class | RCT3 estimate |  | Geometric mean of <br> Time series (since 1990) |
| :--- | :--- | :--- | :--- |
| 2015 | 2900 |  |  |
| 2016 |  | 2443 |  |
| 2017 |  | 2443 |  |

Table 12.23: Whiting in Subarea 4 and Division 7.d. Short-term forecast inputs.


| 1 | 2442840 | 1.255 | 0.11 | 0 | 0 | 0.122 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | . | 0.83 | 0.92 | 0 | 0 | 0.201 |
| 3 | . | 0.643 | 1 | 0 | 0 | 0.305 |
| 4 | . | 0.644 | 1 | 0 | 0 | 0.372 |
| 5 | . | 0.553 | 1 | 0 | 0 | 0.384 |
| 6 | . | 0.553 | 1 | 0 | 0 | 0.474 |
| 7 | . | 0.553 | 1 | 0 | 0 | 0.511 |
| 8 | . | 0.553 | 1 | 0 | 0 | 0.535 |
| Catch |  |  |  |  |  |  |
| Age | Sel | CWt | DSel | DCWt |  |  |
| 1 | 0.00459 | 0.166 | 0.03202 | 0.115 |  |  |
| 2 | 0.04739 | 0.240 | 0.06132 | 0.175 |  |  |
| 3 | 0.09043 | 0.359 | 0.06172 | 0.236 |  |  |
| 4 | 0.14135 | 0.414 | 0.04956 | 0.275 |  |  |
| 5 | 0.27359 | 0.437 | 0.06074 | 0.270 |  |  |
| 6 | 0.23451 | 0.510 | 0.04246 | 0.326 |  |  |
| 7 | 0.16056 | 0.528 | 0.02069 | 0.431 |  |  |
| 8 | 0.15999 | 0.604 | 0.02112 | 0.301 |  |  |
| IBC |  |  |  |  |  |  |
| Age | Sel | CWt |  |  |  |  |
| 1 | 0.00085 | 0.122 |  |  |  |  |
| 2 | 0.0051 | 0.201 |  |  |  |  |
| 3 | 0.00888 | 0.305 |  |  |  |  |
| 4 | 0.01319 | 0.372 |  |  |  |  |
| 5 | 0.02546 | 0.384 |  |  |  |  |
| 6 | 0.0213 | 0.474 |  |  |  |  |
| 7 | 0.01446 | 0.511 |  |  |  |  |
| 8 | 0.01459 | 0.538 |  |  |  |  |
|  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| 2018 |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt |
| 1 | 2442840 | 1.255 | 0.11 | 0 | 0 | 0.122 |
| 2 | . | 0.830 | 0.92 | 0 | 0 | 0.201 |
| 3 | . | 0.643 | 1 | 0 | 0 | 0.305 |
| 4 | . | 0.644 | 1 | 0 | 0 | 0.372 |
| 5 | . | 0.553 | 1 | 0 | 0 | 0.384 |
| 6 | . | 0.553 | 1 | 0 | 0 | 0.474 |
| 7 | . | 0.553 | 1 | 0 | 0 | 0.511 |
| 8 | . | 0.553 | 1 | 0 | 0 | 0.535 |
|  |  |  |  |  |  |  |
| Catch |  |  |  |  |  |  |
| Age | Sel | CWt | DSel | DCWt |  |  |
| 1 | 0.00459 | 0.166 | 0.03202 | 0.115 |  |  |


| 2 | 0.04739 | 0.240 | 0.06132 | 0.175 |
| :--- | :--- | :--- | :--- | :--- |
| 3 | 0.09043 | 0.359 | 0.06172 | 0.236 |
| 4 | 0.14135 | 0.414 | 0.04956 | 0.275 |
| 5 | 0.27359 | 0.437 | 0.06074 | 0.270 |
| 6 | 0.23451 | 0.510 | 0.04246 | 0.326 |
| 7 | 0.16056 | 0.528 | 0.02069 | 0.431 |
| 8 | 0.15999 | 0.604 | 0.02112 | 0.301 |
| IBC |  |  |  |  |
| Age | Sel |  |  |  |
| 1 | 0.00085 | 0.122 |  |  |
| 2 | 0.0051 | 0.201 |  |  |
| 3 | 0.00888 | 0.305 |  |  |
| 4 | 0.01319 | 0.372 |  |  |
| 5 | 0.02546 | 0.384 |  |  |
| 6 | 0.0213 | 0.474 |  |  |
| 7 | 0.01446 |  |  |  |
| Input units are thousands and kg - output in tonnes |  |  |  |  |

## Table 12.23: Whiting in Subarea 4 and Division 7.d. MFDP output table for short-term forecasts.

MFDP version 1a; Run: run2. Time and date: 09:59 02/05/2016, Fbar age range: 2-6; 2015 landings: total 17 706; 4: 13 608 (0.769); 7.d: 4098 (0.231)

| 2016 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch |  |  | Landings |  |  |  | Discards |  | IBC |  |  |  |  |  |  |
| Biomass | SSB | FMult | FBar | Yield | FBar | 4+7.d yield | 4 yield | 7.d yield | FBar | Yield | FMult | FBar | Yield |  | 0.75*Fbar | 1.25*Fbar |  |
| 591174 | 266998 | 1 | 0.2275 | 33601 | 0.1575 | 18537 | 14247 | 4290 | 0.0552 | 13424 | 1 | 0.0148 | 1640 |  | 0.171 | 0.284375 |  |
| 2017 |  |  |  |  |  |  |  |  |  |  |  |  |  | 2018 |  | 2016 TAC 4 | 13678 |
|  |  | Catch |  |  | Landings |  |  |  | Discards |  | IBC | Landings |  |  |  |  |  |
| Biomass | SSB | FMult | FBar | Yield | FBar | 4+7.d yield | 4 yield | 7.d yield | FBar | Yield | FMult | FBar | Yield | Biomass | SSB | 4 TAC change | SSB change |
| 588412 | 310363 | 0 | 0.015 | 1887 | 0.000 | 0 | 0 | 0 | 0.000 | 0 | 1 | 0.015 | 1887 | 622258 | 345826 | -100\% | 11\% No HC fishery |
| . | 310363 | 0.1 | 0.036 | 5481 | 0.016 | 2128 | 1635 | 493 | 0.006 | 1479 | 1 | 0.015 | 1874 | 619163 | 342772 | -88\% | 10\% |
| . | 310363 | 0.2 | 0.057 | 9028 | 0.032 | 4224 | 3246 | 978 | 0.011 | 2944 | 1 | 0.015 | 1860 | 616111 | 339761 | -76\% | 9\% |
| . | 310363 | 0.3 | 0.079 | 12534 | 0.047 | 6290 | 4834 | 1456 | 0.017 | 4397 | 1 | 0.015 | 1847 | 613101 | 336791 | -65\% | 9\% |
| . | 310363 | 0.4 | 0.100 | 15996 | 0.063 | 8325 | 6398 | 1927 | 0.022 | 5837 | 1 | 0.015 | 1834 | 610132 | 333863 | -53\% | 8\% |
| . | 310363 | 0.5 | 0.121 | 19415 | 0.079 | 10330 | 7939 | 2391 | 0.028 | 7264 | 1 | 0.015 | 1821 | 607204 | 330975 | -42\% | 7\% |
| . | 310363 | 0.6 | 0.142 | 22791 | 0.095 | 12305 | 9457 | 2848 | 0.033 | 8678 | 1 | 0.015 | 1808 | 604316 | 328127 | -31\% | 6\% |
| - | 310363 | 0.7 | 0.164 | 26129 | 0.110 | 14252 | 10953 | 3299 | 0.039 | 10081 | 1 | 0.015 | 1796 | 601466 | 325317 | -20\% | 5\% |
| - | 310363 | 0.8 | 0.185 | 29425 | 0.126 | 16171 | 12428 | 3743 | 0.044 | 11471 | 1 | 0.015 | 1783 | 598655 | 322546 | -9\% | 4\% |
| . | 310363 | 0.9 | 0.206 | 32683 | 0.142 | 18062 | 13882 | 4180 | 0.050 | 12850 | 1 | 0.015 | 1771 | 595882 | 319812 | 1\% | 3\% |
| . | 310363 | 1 | 0.228 | 35900 | 0.158 | 19926 | 15314 | 4612 | 0.055 | 14216 | 1 | 0.015 | 1758 | 593145 | 317115 | 12\% | 2\% Fsq |
| . | 310363 | 1.1 | 0.249 | 39080 | 0.173 | 21763 | 16726 | 5037 | 0.061 | 15571 | 1 | 0.015 | 1746 | 590445 | 314455 | 22\% | 1\% |
| . | 310363 | 1.2 | 0.270 | 42223 | 0.189 | 23574 | 18118 | 5456 | 0.066 | 16915 | 1 | 0.015 | 1734 | 587781 | 311830 | 32\% | 0\% |
| . | 310363 | 1.3 | 0.291 | 45329 | 0.205 | 25359 | 19490 | 5869 | 0.072 | 18247 | 1 | 0.015 | 1723 | 585152 | 309240 | 42\% | 0\% |
| . | 310363 | 1.4 | 0.312 | 48397 | 0.220 | 27119 | 20842 | 6277 | 0.077 | 19567 | 1 | 0.015 | 1711 | 582558 | 306685 | 52\% | -1\% |
| . | 310363 | 1.5 | 0.334 | 51431 | 0.236 | 28854 | 22176 | 6678 | 0.083 | 20877 | 1 | 0.015 | 1700 | 579998 | 304163 | 62\% | -2\% |
| $\stackrel{ }{ }$ | 310363 | 1.6 | 0.355 | 54428 | 0.252 | 30564 | 23490 | 7074 | 0.088 | 22176 | 1 | 0.015 | 1688 | 577471 | 301675 | 72\% | -3\% |
| . | 310363 | 1.7 | 0.376 | 57392 | 0.268 | 32251 | 24787 | 7464 | 0.094 | 23464 | 1 | 0.015 | 1677 | 574977 | 299220 | 81\% | -4\% |


| . | 310363 | 1.8 | 0.398 | 60321 | 0.283 | 33914 | 26065 | 7849 | 0.099 | 24741 | 1 | 0.015 | 1666 | 572515 | 296796 | 91\% | -4\% |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| . | 310363 | 1.9 | 0.419 | 63218 | 0.299 | 35555 | 27326 | 8229 | 0.105 | 26008 | 1 | 0.015 | 1655 | 570085 | 294405 | 100\% | -5\% |  |
| . | 310363 | 2 | 0.440 | 66080 | 0.315 | 37172 | 28569 | 8603 | 0.110 | 27264 | 1 | 0.015 | 1644 | 567686 | 292044 | 109\% | -6\% |  |
| . | 310363 | 0.75 | 0.174 | 27184 | 0.118 | 14796 | 11371 | 3424 | 0.041 | 10596 | 1 | 0.015 | 1791 | 600627 | 324495 | -17\% | 5\% | 0.75 * Fsq |
| . | 310363 | 0.64 | 0.150 | 23527 | 0.100 | 12679 | 9744 | 2935 | 0.035 | 9042 | 1 | 0.015 | 1805 | 603735 | 327559 | -29\% | 5.5\% | Fmsy |
| . | 310363 | 1.25 | 0.281 | 43213 | 0.197 | 24075 | 18503 | 5572 | 0.069 | 17408 | 1 | 0.015 | 1731 | 587002 | 311068 | 35\% | 0\% | 1.25 * Fsq |
| . | 310363 | 0.77 | 0.178 | 27759 | 0.121 | 15128 | 11626 | 3501 | 0.042 | 10842 | 1 | 0.015 | 1789 | 600139 | 324015 | -15\% | 4\% | 15\% TAC decrease (4) |
| . | 310363 | 1.06 | 0.239 | 36980 | 0.166 | 20467 | 15730 | 4737 | 0.058 | 14759 | 1 | 0.015 | 1754 | 592300 | 316289 | 15\% | 2\% | 15\% TAC increase (4) |
| . | 310363 | 0.91 | 0.209 | 32369 | 0.144 | 17797 | 13678 | 4119 | 0.050 | 12800 | 1 | 0.015 | 1772 | 596219 | 320152 | 0\% | 3\% | Rollover TAC |
| . | 310363 | 1.00 | 0.227 | 35137 | 0.157 | 19399 | 14909 | 4490 | 0.055 | 13976 | 1 | 0.015 | 1761 | 593867 | 317834 | 9\% | 2\% | Fsq |
| . | 310363 | 1.25 | 0.280 | 43128 | 0.196 | 24025 | 18465 | 5561 | 0.069 | 17372 | 1 | 0.015 | 1731 | 587075 | 311140 | 35\% | 0\% | Fpa |
| . | 310363 | 1.76 | 0.390 | 59714 | 0.278 | 33626 | 25843 | 7783 | 0.097 | 24420 | 1 | 0.015 | 1668 | 572977 | 297247 | 89\% | -4\% | Flim |
| - | 310363 | 3.82 | 0.828 | 125742 | 0.602 | 71846 | 55218 | 16629 | 0.211 | 52478 | 1 | 0.015 | 1418 | 516856 | 241837 | 304\% | -22\% | Bpa, MSY Btrigger |
| $\cdot$ | 310363 | 6.39 | 1.374 | 208120 | 1.007 | 119531 | 91866 | 27665 | 0.353 | 87484 | 1 | 0.015 | 1106 | 446838 | 172741 | 572\% | -44\% | Blim |

Output units in tonnes

Table 24. Frequency of assessment

| stock | Life <br> span | Fsq> = 1.1Fmsy | $\begin{gathered} \text { SSB(2016) } \\ >=M S Y \\ \text { Btrigger } \end{gathered}$ | conribution age 1 to total catches <25\% | $\begin{aligned} & \text { F under- } \\ & \text { estimated } \\ & (<20 \%) \end{aligned}$ | qualifies |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Whg47.d | yes | no | yes | no | yes | NO |



Figure 12.1: Whiting in Subarea 4 and Division 7.d. Landings with discards. Metier with industrial bycatch landings (MIS_MIS_0_0_0_IBC, Denmark, orange) generally does not have discards.


Figure 12.2.1: Whiting in Subarea 4 and Division 7.d. Reported landings (in percent, colured bars)) for each sampled and unsampled fleet, along with cumulative landings (in percent, black line) for fleets in descending order of yield.


Figure 12.2.2: Whiting in Subarea 4 and Division 7.d. Reported discards (in tonnes, colured bars)) for each sampled and unsampled fleet, along with cumulative discards (black line) for fleets in descending order of yield.


Figure 12.3: Whiting in Subarea 4 and Division 7.d. Yield by catch component.


Figure 12.4: Whiting in Subarea 4 and Division 7.d. Proportion of total catch discarded, by age and year.


Figure 12.5: Whiting in Subarea 4 and Division 7.d. Mean weights-at-age (kg) by catch component (black lines) and LOESS smoothers through each time-series of mean weights-at-age (red dashed lines). Catch mean weights are used as stock mean weights.


Figure 12.6: Whiting in Subarea 4 and Division 7.d. Natural mortality estimates from the 2011 SMS key run, used in this year's assessment.


Figure 12.7: Whiting in Subarea 4 and Division 7.d. Survey distribution maps for Ages 1-3+ Q1 2011-2016. Size of the bubbles indicate numbers caught per 30 minutes for each age (on a $\log 10$ scale). The maps are based on the IBTS-Q1 survey in the North Sea.


Figure 12.8: Whiting in Subarea 4 and Division 7.d. Survey distribution maps for ages 0-3+ Q3 20112015. Size of the bubbles indicate numbers caught per 30 minutes for each age (on a $\log 10$ scale). The maps are based on the IBTS-Q3 survey in the North Sea.


Figure 12.9: Whiting in Subarea 4 and Division 7.d. Survey log CPUE (catch per unit effort) at age.

IBTS-Q1


IBTS-Q3


Figure 12.10: Whiting in Subarea 4 and Division 7.d. Log survey indices by cohort for each of the two surveys. The spawning year for each cohort is indicated at the start of each line.


Figure 12.11: Within-survey correlations for the IBTS Q1 survey series, comparing index values at different ages for the same year-classes (cohorts). In each plot, the straight line is a normal linear model fit: a thick line (with black points) represents a significant ( $\mathrm{p}<0.05$ ) regression, while a thin line (with blue points) is not significant. Approximate $95 \%$ confidence intervals for each fit are also shown.


IBTS-Q3

Figure 12.12: Within-survey correlations for the IBTS Q3 survey series, comparing index values at different ages for the same year-classes (cohorts). In each plot, the straight line is a normal linear model fit: a thick line (with black points) represents a significant ( $\mathrm{p}<0.05$ ) regression, while a thin line (with blue points) is not significant. Approximate $95 \%$ confidence intervals for each fit are also shown.

IBTS-Q1


IBTS-Q3


Figure 12.13: Whiting in Subarea 4 and Division 7.d. Survey $\log$ CPUE (catch per unit effort) for the IBTS Q1 and Q3 surveys, by cohort. Each line shows the $\log$ CPUE for the age indicated at the start of the line.


Figure 12.14: Whiting in Subarea 4 and Division 7.d. Summary plots from an exploratory SURBAR assessment, using both available surveys (IBTS Q1 and Q3). Mean mortality Z (ages 2 to 4), relative spawning stock biomass (SSB), relative total biomass (TSB), and relative recruitment. Shaded grey areas correspond to the $\mathbf{9 0} \% \mathrm{CI}$. Green points give the model estimates, while red crosses and black lines give (respectively) the mean and median values from the uncertainty estimation bootstrap.


Figure 12.15: Whiting in Subarea 4 and Division 7.d. Log survey residuals from the SURBAR analysis. Ages are colour-coded, and a LOESS smoother (span $=2$ ) has been fitted through each age time-series.


Figure 12.16: Whiting in Subarea 4 and Division 7.d. Parameter estimates from SURBAR analysis. Top row: age, year and cohort effect estimates as box-and-whisker plots. Bottom row: estimates as line plots with $\mathbf{9 0} \%$ confidence intervals.

## Commercial Catch Data



Figure 12.17: Whiting in Subarea 4 and Division 7.d. Log catch curves by cohort for total catches.


Figure 12.18: Whiting in Subarea 4 and Division 7.d. Negative gradients of log catches per cohort, averaged over ages 2-6. The $x$-axis represents the spawning year of each cohort.


Catch numbers at age

Figure 12.19: Whiting in Subarea 4 and Division 7.d. Correlations in the catch-at-age matrix (including the plus-group for ages 8 and older), comparing estimates at different ages for the same yearclasses (cohorts). In each plot, the straight line is a normal linear model fit: a thick line (and black points) represents a significant ( $\mathbf{p}<0.05$ ) regression, while a thin line (and blue points) is not significant. Approximate $95 \%$ confidence intervals for each fit are also shown.


Figure 12.20: Whiting in Subarea 4 and Division 7.d. Stock summary plots for single-fleet XSA runs. Only the more recent segments of the EngGFS and ScoGFS surveys have been used here. Final year (2015) values of SSB and mean $F(2-4)$ are plotted against each other in the upper right plot.


Figure 12.21: Whiting in Subarea 4 and Division 7.d. Log catchability residuals for single-fleet FLXSA assessments (negative values as black bubbles, positive values as yellow bubbles).


Figure 12.22: Whiting in Subarea 4 and Division 7.d. Comparisons of stock summary estimates from the final XSA (blue) and SURBAR (red) models. To facilitate comparison, values have been meanstandardised using the year range for which estimates are available from all three models. The SURBAR estimates are plotted along with their $\mathbf{9 0} \%$ confidence bounds (shaded pink regions).


Figure 12.23: Whiting in Subarea 4 and Division 7.d. Log catchability residuals for final FLXSA assessment (negative values as black bubbles, positive values as yellow bubbles).


Figure 12.24: Whiting in Subarea 4 and Division 7.d. Contribution to survivors' estimates in final FLXSA assessment.


Figure 12.25: Whiting in Subarea 4 and Division 7.d. Summary plots for final FLXSA assessment.


Figure 12.26.1: Whiting in Subarea 4 and Division 7.d. Retrospective plots for final FLXSA assessment.


Figure 12.26.2: Whiting in Subarea 4 and Division 7.d. Retrospective plots for final FLXSA assessment, relative differences.


Figure 12.27: Whiting in Subarea 4 and Division 7.d. Stock-recruitment plot from final FLXSA assessment.


Figure 12.28: Whiting in Subarea 4 and Division 7.d. FLXSA F at age estimates for 2012-2015, along with scaled mean exploitation used for the forecast.


Figure 12.29: Whiting in Subarea 4 and Division 7.d. Historical assessment comparison plot.

### 12.2 Whiting in Division 3.a

### 12.2.1 General

### 12.2.1.1 Stock definition

There is a paucity of information on the population structure of whiting in Division 3.a (the Skagerrak-Kattegat area). No genetic surveys have been conducted, nor otolith based surveys. Tagging of whiting has previously been undertaken, yet these data need to be re-examined. Results from modelled survey data (SURBAR) are inconclusive regarding independent population dynamics in Division 3.a in comparison with the North Sea, presumably due to the need of age readings in 3.a. The drop in landings in the beginning of the 1990s gives however an indication of local stock structure, as this reduction was not paralleled by any similar event in the North Sea.

### 12.2.1.2 Ecosystem aspect

No new information was presented at the Working Group. A summary of available information on ecosystem aspects is presented in the Stock Annex prepared at ICESWKROUND (2009).

### 12.2.1.3 Fisheries

Information on the fisheries was provided by Sweden in terms of the spatial distribution of the Swedish landings in 2015 using logbooks information. The plot is reported in Figure 12.1.1 and showed that higher landings were taken along the Swedish coastline than in the offshore Skagerrak. A summary of available information on fisheries is presented in the Stock Annex prepared at ICES-WKROUND (2009). Discards estimates are available since 2003. Information on derivation of discards is presented in the Stock Annex.

### 12.2.1.4Data available

According to the WKLIFE categorisation of various levels of available data for assessment, whiting in Division 3.a can be considered to be a stock for which survey based indices are available, indicating trends. This survey data have been used for an exploratory assessment.

Total landings are shown in Table 12.1.1.
The WGNSSK in 2015 used IBTS indices per area (Skagerrak and Kattegat) and BITS indices (Kattegat) for plotting CPUE per quarter of fish of total length $>21 \mathrm{~cm}$, which corresponds to the $50 \%$ point of the maturity ogive of whiting in the North Sea. Plots of the IBTS Q1 and IBTS Q3 per area are shown in Figures 12.1.2 and BITS Q1 and Q4 in Figure 12.1.3. IBTS Q3 indicate high inter-annual variability in recruitment. IBTS Q1 in Kattegat shows a marked increase in CPUE in 2015. This is assigned to one single haul dominating the data series. Survey abundance indices are plotted in log-mean standardised form by year and cohort in Figure 12.1.4a for the IBTS Q1 survey, together with log-abundance curves and associated negative gradients for the age range $2-4$. Similar plots are shown for the IBTS Q3 survey in Figure 12.1.4b. Year effects occur (top left) and the importance of cohorts fluctuate through the time-series (top right) indicating migratory behaviour. No clear pattern of total mortality (bottom right).

### 12.2.2 Data analyses

### 12.2.2.1 Exploratory survey-based analysis

Based on the information provided by the IBTS mean age indices for Q1 and Q3 a SURBAR analysis was performed. The summary plot from this run is given is Figure 12.1.5 and indicated great uncertainties in all parameter values of relative spawning stock biomass (SSB), relative total biomass (TSB) and mean mortality (Z) with highly erratic patterns.
The log index values (number at age) plotted against numbers at age +1 of the same cohort in the following year are shown in Figure 12.1.6. For both IBTS Q1 and IBTS Q3 surveys the different plots indicated that internal consistency was virtually absent, impeding cohort analysis in the stock for the present. Log residual estimates per age class for IBTS Q1 and IBTS Q3 are shown in Figure 12.1.7.

### 12.2.2.2 Conclusions drawn from exploratory analysis

The lack of internal consistency in the available survey indices prevents analytical assessment. This internal inconsistency could be related to a) age reading problems, and/or b) a mixture of several stock components leading to unaccounted migrations. As the survey-based assessment cannot be used as a basis for advice, the stock is thus classified, according to the ICES rules for data limited stocks, as belonging to category 5.2. No new data were presented at the WGNSSK 2016 to change the perception of the stock. WGH 3.a is up for Benchmark in 2017 and the suggested workplan was updated.

### 12.2.2.3 Advice

DLS-category 5.2, which is based catch information only.. Multi-annual advice is given (2016). There are no new data that change the perception of the stock status.

Table 12.1.1 Nominal landings ( $\mathbf{t}$ ) of Whiting from Division 3.a as supplied by the Study Group on Division 3.a Demersal Stocks (ICES 1992b) and updated by the Working Group, and WG estimate of Discards.

| Year |  | Denmark (1) |  | Norway | Sweden | Others | Total | WG estimate of Discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 19,018 |  |  | 57 | 611 | 4 | 19,690 |  |
| 1976 | 17,870 |  |  | 48 | 1,002 | 48 | 18,968 |  |
| 1977 | 18,116 |  |  | 46 | 975 | 41 | 19,178 |  |
| 1978 | 48,102 |  |  | 58 | 899 | 32 | 49,091 |  |
| 1979 | 16,971 |  |  | 63 | 1,033 | 16 | 18,083 |  |
| 1980 | 21,070 |  |  | 65 | 1,516 | 3 | 22,654 |  |
|  | Total consumption | Total industrial | Total |  |  |  |  |  |
| 1981 | 1,027 | 23,915 | 24,942 | 70 | 1,054 | 7 | 26,073 |  |
| 1982 | 1,183 | 39,758 | 40,941 | 40 | 670 | 13 | 41,664 |  |
| 1983 | 1,311 | 23,505 | 24,816 | 48 | 1,061 | 8 | 25,933 |  |
| 1984 | 1,036 | 12,102 | 13,138 | 51 | 1,168 | 60 | 14,417 |  |
| 1985 | 557 | 11,967 | 12,524 | 45 | 654 | 2 | 13,225 |  |
| 1986 | 484 | 11,979 | 12,463 | 64 | 477 | 1 | 13,005 |  |
| 1987 | 443 | 15,880 | 16,323 | 29 | 262 | 43 | 16,657 |  |
| 1988 | 391 | 10,872 | 11,263 | 42 | 435 | 24 | 11,764 |  |
| 1989 | 917 | 11,662 | 12,579 | 29 | 675 | - | 13,283 |  |
| 1990 | 1,016 | 17,829 | 18,845 | 49 | 456 | 73 | 19,423 |  |
| 1991 | 871 | 12,463 | 13,334 | 56 | 527 | 97 | 14,041 |  |
| 1992 | 555 | 3,340 | 3,895 | 66 | 959 | 1 | 4,921 |  |
| 1993 | 261 | 1,987 | 2,248 | 42 | 756 | 1 | 3,047 |  |
| 1994 | 174 | 1,900 | 2,074 | 21 | 440 | 1 | 2,536 |  |
| 1995 | 85 | 2,549 | 2,634 | 24 | 431 | 1 | 3,090 |  |
| 1996 | 55 | 1,235 | 1,290 | 21 | 182 | - | 1,493 |  |
| 1997 | 38 | 264 | 302 | 18 | 94 | - | 414 |  |
| 1998 | 35 | 354 | 389 | 16 | 81 | - | 486 |  |
| 1999 | 37 | 695 | 732 | 15 | 111 | - | 858 |  |
| 2000 | 59 | 777 | 836 | 17 | 138 | 1 | 992 |  |
| 2001 | 61 | 9701 | 1,0311 | 27 | 126 | + | 1,1841 |  |
| 2002 | 101 | 9751 | 1,0761 | 23 | 127 | 1 | 1,2271 |  |
| 2003 | 93 | 6541 | 7471 | 20 | 71.9 | 2 | 840.91 | 429 |
| 2004 | 93 | 1,1201 | 1,2131 | 17 | 74 | 1 | 1,3051 | 909 |
| 2005 | 49 | 9071 | 9561 | 13 | 73 | 0 | 1,0421 | 299 |
| 2006 | 591 | 2901 | 3491 | n/a | 85.92 | n/a | 434.92 | 331 |
| 2007 | 532 | 2782 | 3312 | 14 | 82 | 1 | 4282 | 561 |
| 2008 | 522 | 2882 | 3402 | 14 | 52 | n/a | 4062 | 241 |
| 2009 | 712 | 1732 | 2442 | 10.3 | 33.82 | - | 288.12 | 128 |
| 2010 | 41 | 165 | 206 | 9.7 | 29.7 | - | 245.4 | 291 |
| 2011 | 40 | 44 | 84 | 8.3 | 20.4 | 0.2 | 112.9 | 794 |
| 2012 | 30 | 6.8 | 37 | 15.5 | 9.6 | 0.8 | 62.9 | 277 |
| 2013 | 29 | 102 | 131 | 8.4 | 14.5 | 1.0 | 155 | 591 |
| 2014 | 49 | 346 | 395 | 4.8 | 37.6 | 1.3 | 439 | 579 |
| 2015 | 74 | 572 | 646 | 5.9 | 55.681 | 5.1 | 713.4 | 604 |

${ }^{1}$ Values from 1992 updated by WGNSSK (2007).
${ }^{2}$ Values updated by WGNSSK (2011).


Figure 12.1.1. Whiting in Division 3.a. Spatial distribution of the total landings of whiting 3.a in Swedish fisheries 2014 from logbooks information.

## Skagerrak



Kattegat


Figure 12.1.2. Whiting in Division 3.a (Skagerrak and Kattegat). IBTS CPUE per area Q1 covering the years 1967-2015 and Q3 covering the years 1991-2014.


Figure 12.1.3. Whiting in Division 3.a S (Kattegat). BITS CPUE per Q1 and Q4 covering the years 1992-2015 and 1992-2014, respectively.


Figure 12.1.4a Whiting in Division 3.a. Log mean standardized indices plotted by year (top left) and cohort (top right), log abundance curves (bottom left) and associated negative gradients for each cohort across the reference fishing mortality of age 2-4 (bottom right), for the IBTSQ1 groundfish survey (NS-IBTS Delta-GAM index)


Figure 12.1.4b Whiting in Division 3.a. Log mean standardized indices plotted by year (top left) and cohort (top right), log abundance curves (bottom left) and associated negative gradients for each cohort across the reference fishing mortality of age 2-4 (bottom right), for the IBTSQ3 groundfish survey (NS-IBTS Delta-GAM index)


Figure 12.1.5. Whiting in Division 3.a. SURBAR analysis. Mean mortality $\mathbf{Z}$ (ages 2 to 4), relative spawning stock biomass (SSB), relative total biomass (TSB), and relative recruitment. Shaded grey areas correspond to the $90 \%$ CI. Green points give the model estimates, while red crosses and black lines give (respectively) the mean and median values from the uncertainty estimation bootst

IBTS_Q1


IBTS_Q3


Figure 12.1.6 Whiting in Division 3.a. SURBAR analysis. IBTS indices per age class 1-5 for Q1 covering the years 1980-2015 and Q3 covering the years 1991-2014. The log index values (number at age) plotted against numbers at age +1 of the same cohort in the following year.


Figure 12.1.7 Whiting in Division 3.a. SURBAR analysis. Log residual estimates per age class for IBTS Q1 (upper line plots) and IBTS Q3 (lower line plots).

## 13 Haddock in Subarea 4 and Divisions 3.a. 20 and 6.a (North Sea, Skagerrak and West of Scotland)

Until 2014, haddock in Subarea 4 and Divisions 3.a. 20 and 6.a (referred to hereafter as Northern Shelf haddock) were assessed as two separate stocks: Subarea 4 and Division 3.a. 20 by WGNSSK, and Division 6.a by WGCSE. The 2014 Benchmark Workshop for Northern Haddock Stocks (ICES-WKHAD 2014) concluded that the two notional haddock stocks should be assessed as one stock. This section presents the third annual ICES assessment of Northern Shelf haddock.

During the 2016 WGNSSK meeting, problems were identified with the update haddock assessment. These could not be rectified during the meeting, and a separate Interbenchmark Group (IBPHaddock) was convened to address the issue during the summer of 2016 (ICES-IBPHaddock 2016). IBPHaddock concluded that a) the existing TSA model code contained an error (and had done so since 2015), and that b) the corrected code produced a retrospective bias, which was addressed by modifying the treatment of recent above-average year-classes. The following Section reports the results of the corrected model produced during the IBPHaddock process, which was used as the basis for the 2017 advice (and revised 2016 advice) published in November 2016.

### 13.1 General

### 13.1.1 Ecosystem aspects

Ecosystem aspects are summarised in the Stock Annex.

### 13.1.2 Fisheries

A general description of the fishery (along with its historical development) is presented in the Stock Annex. Most of the information presented below and in the Stock Annex pertains to the Scottish fleet, which takes the largest proportion of the haddock stock. This fleet is not just confined to the Northern Shelf area, as vessels will sometimes operate in Divisions 6.b (Rockall) and 5.b (Faroes).

### 13.1.2.1 Changes in fleet dynamics

There have been no decommissioning schemes affecting haddock fisheries since the major rounds in 2002 and 2004. A number of Scottish vessels have been taking up opportunities for oil support work during recent years with a view to saving quota and days at sea.

With the relatively limited cod and whiting quotas in recent years, many vessels have tended to concentrate more on the haddock fishery, with others taking the opportunity to move between the Nephrops and demersal fisheries (particularly during 2006 and 2007 - there may have been fewer boats changing focus in this way from 2008 to 2015). Accompanying the change in emphasis towards the haddock fishery, there has also been a tendency to target smaller fish in response to market demand. Some trawlers operating in the east of the North Sea have used 130 mm mesh and this is likely to have improved selectivity for haddock. Fish from the 2009 year-class still form the bulk of haddock catches (although the modestly large 2014 year-class started to contribute to landings towards the end of 2015), and discarding rates for the 2009 year-class fish declined during 2012 and 2013 as they grew beyond the minimum landings size. The decline may also have been due to other measures related to the Scottish Conservation

Credits scheme (CCS; see Section 13.1.4). Discard rates in 2015 increased once again, although the reasons for this are not clear.

Specific information on changes in the Scottish fleet during 2011-2015 was not provided to WGNSSK in 2016. It is difficult to reach a firm conclusion on the likely effect of recent fishery changes on haddock mortality. Changes in gear that were required to qualify for the Scottish CCS are likely to have reduced bycatch (and therefore discards) of haddock in the Nephrops fishery in particular. The inclusion of Scottish vessels in the CCS has been mandatory since the beginning of 2009, and compliance has been close to $100 \%$. Cod avoidance under the real-time closures scheme (which is a component of the CCS) could also have moved vessels away from haddock concentrations, but the extent of this depends on how closely cod and haddock distributions are linked, and on how successful the avoidance strategies have been. On the other hand, vessels catching fewer cod may have increased their exploitation of haddock in order to maintain economic viability. It is unclear what changes in fleet dynamics and fishing behaviour may be caused by the impending EU landings obligation, due to be implemented for the majority of fleets catching Northern Shelf haddock in January 2016.

Following trials during 2010-2013, 26 Scottish demersal whitefish vessels participated in the 2014 Fully Documented Fishery (FDF) scheme (although 3 vessels left the scheme during the year). Similar trials have been conducted during various periods by Denmark, England, Germany, Sweden and the Netherlands. In the Scottish North Sea FDF trials, vessels are exempt from some effort restrictions and are allocated additional cod quota: in return, they must carry monitoring cameras and land all cod caught. It is not clear what the impact would be on haddock fisheries of an enforceable discard ban for cod, and in data collation for the haddock assessment it was assumed that FDF vessels would have similar haddock discard patterns as other vessels, but this remains to be verified. It should be noted that the Scottish FDF schemes implemented to date have all been restricted to the North Sea: cod discarding from CCTV vessels has remained legal in Division 6.a, and indeed has been mandatory for over-quota cod. The Scottish FDF scheme for 2015 continued without a break from the end of 2014, and included 24 vessels (although 6 left during the year). Currently, 12 vessels are participating in the scheme in 2016: the uptake of the scheme has declined due to concerns about monitoring of discards under the EU Landing Obligation.

### 13.1.2.2Additional information provided by the fishing industry

Haddock are still the mainstay of the Scottish whitefish fleet, and have become increasingly so following cod-avoidance initiatives under the Scottish Conservation Credits scheme.

### 13.1.3 ICES advice

### 13.1.3.1 3.a. 20 ICES advice for 2015

### 13.1.3.1.1 Subarea 4 and Divisions 3.a. 20 and 6.a

From 2014 onwards, ICES advice is provided on the basis of the assessment stock unit, rather than the component management units. Hence, in November 2014 (following the application of the AGCREFA (ICES-AGCREFA 2008) update protocol, ICES concluded the following for the Northern Shelf haddock stock:

ICES advises on the basis of the MSY approach that catches should be no more than $68690 t$ for the whole assessment area. If rates of discards and industrial bycatch do not change from the average of the last three years (2011-2013), this implies human
consumption landings of no more than $50163 t$. Measures to reduce discards should be taken in order to protect the incoming recruitment.

### 13.1.3.2 ICES advice for 2016

### 13.1.3.2.1 Subarea 4 and Divisions 3.a. 20 and 6.a

In June 2015, ICES concluded the following:
ICES advises that when the MSY approach is applied, catches in 2016 should be no more than 74854 tonnes. If this stock is not under the EU landing obligation in 2016 and discard rates do not change from the average (2012-2014), this implies landings of no more than 61930 tonnes.
The application of the update protocol in September 2015 did not lead to a revision of the advice for this stock.

### 13.1.4 Management

Until 2014, North Sea haddock (Subarea 4 and Division 3.a.20) were jointly managed by the EU and Norway under an agreed management plan, the details of which are given in the Stock Annex. However, the validity and sustainability of the management plan when applied to the wider Northern Shelf area had not been evaluated by ICES, and advice could not be provided on the basis of the plan as a consequence. A separate management plan for Division 6.a was evaluated by ICES in 2008 to be precautionary, but similarly cannot be used to provide advice for the full stock area. A management plan for Northern Shelf haddock was to have been developed during 2015, but this has not yet occurred as the basis for management of shared EU-Norway stocks has still to be agreed. In the meantime the stock is managed according to advice based on the ICES MSY approach.

During 2008, 15 real-time closures (RTCs) were implemented under the Scottish Conservation Credits Scheme (CCS). In 2009, 144 RTCs were implemented, and the CCS was adopted by 439 Scottish and around 30 English and Welsh vessels. In 2010 there were 165 closures, and from July 2010 the area of each closure increased (from 50 square nautical miles to 225 square nautical miles). In more recent years, the following numbers of closures were implemented: 185 (2011), 173 (2012), 166 (2013), 94 (2014) and 97 (2015). 114 closures were implanted during 2016, although the scheme was suspended on $20^{\text {th }}$ November and there are no plans for its reintroduction. The CCS had two central themes aimed at reducing the capture of cod through (i) avoiding areas with elevated abundances of cod through the use of Real Time Closures (RTCs) and (ii) the use of more species selective gears. Within the scheme, efforts were also being made to reduce discards generally. Although the scheme was intended to reduce mortality on cod, it undoubtedly had an effect on the mortality of associated species such as haddock.

Studies tracking Scottish vessels during 2009-2010 concluded that vessels did indeed move from areas of higher to lower cod concentration following real-time closures during the first and third quarters, although there was no significant effect during the second and fourth quarters; see Needle and Catarino (2011). In a subsequent analysis, Needle (2012) showed that the net effect of RTCs appeared to be to attract vessels, although the movement towards RTCs may have been coincidental. However, the effect of these changes in behaviour on the haddock stock is still under investigation.

In 2015, 24 Scottish demersal whitefish vessels participated in a trial Fully Documented Fishery (FDF) scheme, following similar schemes during 2010-2014. The Scottish FDF scheme for 2016 began in April, and is being run along similar lines to previous years. At the time of writing, in May 2016, 12 vessels were included in the scheme, with each receiving an increase in cod quota and extra days at sea in return for landing all cod caught.

In early 2008, a one-net rule was introduced in Scotland as part of the CCS. This is likely to have improved the accuracy of reporting of landings to the correct mesh size range. The remaining technical conservation measures in place for the haddock fisheries in Subarea 4 and Division 3.a. 20 and 6.a are summarised in the Stock Annex.

Annual management of the fishery operates through TACs for three discrete areas. The first is Subarea 4 (and EU Waters of 2.a). The 2015 and 2016 TACs for haddock in this area were 40711 t and 61933 t respectively. The second is Division 3.a (EU waters), for which the TACs for 2015 and 2016 were 2504 t and 3926 t respectively. The third is Division 6.a, for which the TACs in 2015 and 2016 were 4536 t and 6462 t respectively.

### 13.2 Data available

### 13.2.1 Catch

Official landings data for each country participating in the fishery are presented in Table 13.2.1, together with the corresponding WG estimates and the agreed international quota (listed as "total allowable catch" or TAC). Since 2012, international data on landings and discards have been collated through the InterCatch system (see Section 1.2). Figure 13.2.1 and Tables 13.2.2 to 13.2.4 summarise the proportion of landings in the combined Northern Shelf area, for which samples have been provided. While there are a large number of fleets for which landings have not been sampled, the overall contribution of these fleets to total landings is small and more than $90 \%$ of landings by weight have been sampled appropriately. Age compositions for the remaining landings have therefore been determined by averaging across the available sampling (as for last year), without consideration of quarter, country or gear type. Similarly, discard observations are available for the fleets landing the vast majority of haddock (see Figure 13.2.2), so discard rates for the remaining fleets have also been inferred using simple averaging.
The full time series of landings, discards and industrial by-catch (IBC) is presented in Table 13.2.5. These data are illustrated further in Figure 13.2.3. The total landed yield of the international fishery has been relatively stable since 2007. The WG estimates (Table 13.2.5) suggest that haddock discarding (as a proportion of the total catch) decreased significantly during 2013, and the discard rate for that year was the lowest in the time series at $7.2 \%$ by weight. This may have been due in part to fleet behaviour changes related to cod avoidance measures, but also to the weak year-classes since 2009 (implying that the bulk of the catch was large, mature fish that are less likely to be discarded). The discard rate increased once more to around $11 \%$ by weight in 2014 and around $15 \%$ in 2015, although the reasons for this are not known. The recent changes in discarding are not consistent across ages (Figure 13.2.4).

Subarea 4 discard estimates are derived from data submitted by Denmark, Germany, England and Scotland. As Scotland is the principal haddock fishing nation in that area, Scottish discard practices dominate the overall estimates. DCF regulations oblige only the UK (Scotland and England) and Denmark to submit discard age-composition data for Subarea 4. Division 3.a discard estimates are derived from data submitted by Den-
mark. Division 6.a discard estimates are provided by UK (Scotland) and Ireland. Industrial bycatch (IBC) has declined considerably from the high levels observed until the late 1970s.

Estimated discard rates can be calculated using video data from Scottish vessels carrying cameras (as part of the FDF scheme described in Section 13.1.2). Neither fish ages nor weights can be measured directly using video, but a method has been developed in Scotland for estimating discard rates by measuring numbers and lengths of discarded fish and applying existing weight-length relationships to obtain a discarded weight, which can then be compared with the total landed weight (see Needle et al 2015). The lack of age information currently impedes the use of these estimates in the ICES assessment process, but work is underway in Scotland and elsewhere to address this.

### 13.2.2 Age compositions

Total catch-at-age data are given in Table 13.2.6, while catch-at-age data for each catch component are given in Tables 13.2.7 to 13.2.9. The fishery in 2015 (landings for human consumption) was still strongly reliant on the 2009 and 2012 year-classes. In the past, vessels have very seldom exhausted their quota in this fishery, and previous discarding behaviour is thought to be driven by a complicated mix of economic and other marketdriven factors. From 2016 onwards, haddock fishing is covered by the EU Landing Obligation.

### 13.2.3 Weight at age

Weight-at-age for the total catch in the North Sea is given in Table 13.2.10. Weight-atage in the total catch is a number-weighted average of weight-at-age in the human consumption landings, discards and industrial bycatch components. Weight-at-age in the stock is assumed to be the same as weight-at-age in the total catch. The mean weights-at-age for the separate catch components are given in Tables 13.2.11 to 13.2.13 and are illustrated in Figure 13.2.5: this shows the declining trend in weights-at-age for older ages in total catch and landings, as well as increasing trends for younger ages and some evidence for reduced growth rates for large year classes. Jaworski (2011) concluded that linear cohort-based growth models are the most appropriate method for characterising haddock growth, and these are used in the short-term forecast (Section 13.6).

### 13.2.4 Maturity and natural mortality

Maturity is assumed to be fixed over time and knife-edged at age 3 (that is, all fish aged $0-2$ are assumed to be immature, all fish aged 3 and older are assumed to be fully mature). Natural mortality varies with age and year as shown in Figure 13.2.6 and Table 13.2.14. The general basis for these estimates is described in the Stock Annex, and these values shown here are derived from the WGSAM 2014 key run (as revised in 2015).

### 13.2.5 Catch, effort and research vessel data

The survey data available are summarised in the following table: data used in the final assessment are highlighted in bold.

| Area | Country | Quarter | Code | Year range | Age <br> RANGE |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Subarea 4 | Scotland | Q3 | ScoGFS Aberdeen Q3 | 1982-1997 | 0-8 |
| Subarea 4 | Scotland | Q3 | ScoGFS Q3 GOV | 1998present | 0-8 |
| Subarea 4 | England | Q3 | EngGFS Q3 GRT | 1977-1991 | 0-9 |
| Subarea 4 | England | Q3 | EngGFS Q3 GOV | 1992- <br> present | 0-9 |
| Subarea 4 and <br> Division 3.a | International | Q1 | IBTS Q1 | 1983present | 1-5 |
| Subarea 4 and <br> Division 3.a | International | Q3 | IBTS Q3 | 1991- <br> present | 0-5 |
| Subarea 6.a | Scotland | Q1 | ScoGFS-WIBTS Q1 | 1985-2010 | 1-8 |
| Subarea 6.a | Scotland | Q1 | New ScoGFSWIBTS Q1 | 2011present | 1-8 |
| Subarea 6.a | Scotland | Q4 | ScoGFS-WIBTS Q4 | 1996-2009 | 0-7 |
| Subarea 6.a | Scotland | Q4 | New ScoGFSWIBTS Q4 | 2011present | 0-7 |
| Subarea 6.a | Ireland | Q4 | IGFS-WIBTS-Q4 | 1993-2002 | 0-8 |
| Subarea 6.a | Ireland | Q4 | New IGFS-WIBTSQ4 | 2003- <br> present | 0-8 |

The 2014 benchmark meeting (ICES-WKHAD 2014) concluded that only the North Sea IBTS Q1 and Q3 survey indices should be used to tune the Northern Shelf assessment. The West of Scotland surveys conducted by Scotland and Ireland covered too small a proportion of the overall stock area to be considered reliable indicators of overall stock dynamics, and the separate English and Scottish North Sea indices were only used previously because of the historical timing of the working group (WGNSSK met in early October when IBTS Q3 was not yet available). ICES-WKHAD (2014) recommended that the IBTS working group consider whether the North Sea IBTS Q1 and West of Scotland ScoGFS Q1 indices could be combined, but this is for future consideration.

Data used for the calibration of the assessment are presented in Table 13.2.15. Surveybased abundance distributions by age and year are given in Figures 13.2.7 (North Sea IBTS Q1), 13.2.8 (North Sea IBTS Q3) and 13.2.9 (Scottish West Coast IBTS Q4)). These demonstrate the concentration of North Sea haddock towards the north and west of the North Sea, quite widely along the continental shelf to the west of Scotland. The modestly large 2014 year-class is evident in all three surveys. Abundance trends in survey indices are shown in Figure 13.2.10. These indicate reasonably good consistency in stock signals from the two North Sea surveys, and support the perception of a modestly large 2014 year-class.

### 13.3 Data analyses

The assessment has been carried out using TSA (Fryer 2002) as the main assessment method. The results of SURBAR and SAM analyses are also shown, to corroborate (or otherwise) the main assessment.

### 13.3.1 Exploratory catch-at-age-based analyses

The catch-at-age data, in the form of log-catch curves linked by cohort (Figure 13.3.1), indicates partial recruitment to the fishery for most cohorts up to age 2. Gradients between consecutive values within a cohort have reduced considerably for some recent cohorts, reflecting a reduction in fishing mortality, although catch curves are considerably more variable in recent years suggesting less consistent catch data (which may reflect the lower sample size available from reduced landings). Figure 13.3 .2 plots the negative gradient of straight lines fitted to each cohort over the age range $2-4$, which can be viewed as a rough proxy for average total mortality for ages $2-4$ in the cohort. These negative gradients are also lower in most recent cohorts, and the negative gradient measure for the 2010 cohort is the lowest in the time-series: it is itself negative, which in the absence of other information would indicate that the 2010 was increasing in size over time. As this cannot be the case, it suggests potential problems with recent catch data. It can also be seen that the negative gradient for the 2010 cohort (from ages $2-4$ ) rises sharply, which suggests that fishing mortality may have increased in the most recent time-period.

Cohort correlations in the catch-at-age matrix (plotted as log-numbers) are shown in Figure 13.3.3. These correlations show good consistency within cohorts up to the plusgroup, verifying the ability of the catch-at-age data over the full time-series to track relative cohort strengths (although data for ages 0 and 1 are slightly more variable, and recent years may be problematic as discussed above).

An exploratory SAM assessment was conducted, using the run settings stipulated in ICES-WKHAD (2014). The stock summary and residual plots from this run are given in Figure 13.3.4. The SAM assessment follows similar trends to the final TSA assessment, although the F estimates are less variable (see also Figure 13.3.10). There is evidence of some retrospective underestimation of mean F in the SAM runs, with a corresponding retrospective overestimation of SSB.

### 13.3.2 Exploratory survey-based analyses

A SURBAR run (ICES-WKADSAM 2010, Needle 2015) was carried out using the same combination of tuning indices as the TSA and SAM assessments. The summary plot from this run is given is Figure 13.3.5, which indicates good precision in relative trend estimates for mortality, biomass and recruitment. The SURBAR residual plot in Figure 13.3.6 shows that the surveys agree more closely in recent years than was the case at the 2014 WGNSSK meeting, although there remains an indication of some conflict (mostly negative residuals for Q1 and a more even spread for Q3. The plot of survey catch curves also shows reasonable consistency (Figure 13.3.7). The plots of meanstandardised log survey indices by age and cohort (Figure 13.3.8) and the pairwise within-survey correlations (Figure 13.3.9) show that both surveys track year-class strength well through the population overall. The results are discussed further in Section 13.3.4 below.

### 13.3.3 Assessment problems (May 2016)

During the May 2016 WGNSSK meeting, the update TSA assessment methodology was applied to the available catch and survey data, as stipulated in the Stock Annex. However, analysis of retrospective TSA runs revealed that, unlike for previous FLXSA, SAM and SURBAR retrospective analyses from previous WGs for North Sea haddock, and the 2014 and 2015 TSA assessments for Northern Shelf haddock, there was significant
retrospective bias in this year's assessment. This bias caused an annual upwards adjustment in fishing mortality and a corresponding reduction in SSB. During the May meeting, WGNSSK hypothesised that this bias was caused by the way in which TSA partitions the variance in F estimates between transitory and persistent effects, but could not identify the precise cause with any certainty. The final assessment was also significantly different to the final assessment presented last year, with much lower SSB estimates and much higher F estimates in many years. The combination of internal problems (retrospective bias) and lack of consistency with last year's assessment led WGNSSK to recommend that the final assessment produced in May not be used as the basis for advice. This suggestion was subsequently confirmed by the Advice Drafting Group for the North Sea (ADGNS), which convened in June. A Review Group also met, which rejected the XSA model that had been proposed as a stop gap remedy, and the assessment was finally rejected. ICES then initiated an Interbenchmark Procedure for Haddock (IBPHaddock), which met by correspondence during the summer of 2016 to address the problems with the assessment.

### 13.3.4 Assessment revisions from IBPHaddock (November 2016)

The report of the IBP (ICES-IBPHaddock 2016) provides detailed information on the analyses performed and conclusions reached. The Executive Summary reads as follows:

The Interbenchmark Workshop on Haddock in Subarea 4, Division 6.a and Subdivision 3.a. 20 (IBPHaddock), chaired by José De Oliveira (UK) took place by correspondence during 4 meetings spread over several weeks (29 June - 29 September 2016). There were eight participants, including two external reviewers (both from the USA) and scientists from the UK and Germany. The main focus of the IBP was to investigate the cause of the apparent failure of the TSA model, to remedy this failure, if possible, or to consider alternative models, if not, and to re-estimate reference points based on the newly selected model. The IBP identified the problem as a retrospective pattern caused by the way in which the larger post-1999 recruitment events were treated, and was able to find a TSA model configuration that remedied this problem; this was achieved by not treating any of the post-1999 year-classes as "outstanding". The post-1999 period was then used as a basis for estimating reference points, apart from Blim which was taken to be the lowest SSB that produced an outstanding year-class (1979).

In addition to a revision of the way that post-1999 larger year-classes were treated, the new configuration also limited the assumption of flat-topped selectivity to ages 7 and 8+ (thus allowing for dome-shaped selectivity in recent years, as indicated by exploratory SAM and XSA runs), and used external (data-driven) estimates for the CVs of the landings and discards data (rather than simpler assumptions used in previous implementations).

The outcome of the analysis was a newly-configured TSA model which presented greatly reduced retrospective bias, when compared to the update TSA model produced at the May 2016 WGNSSK meeting. This was used in turn to generate revised reference points (see Section 13.8) and an updated short-term forecast, which in addition incorporated the available 2016 Q3 IBTS survey data (thus removing the need for the usual update protocol).

During the IBP, it also became apparent that an error had arisen in the TSA code developed for the assessment presented by the 2015 WG . The error was related to the
indexing of matrices and was very difficult to discover - it also only affected assessments which included age-0 data (and thus did not impinge on the TSA assessments used for other ICES stocks). The effect of the error was to overestimate SSB and underestimate $F$, leading to a falsely optimistic impression of the stock. ICES accordingly issued revised advice for the 2016 fishing year in November 2016, replacing the previous F (msy)-based advice of a $30 \%$ catch increase with advice for a $2 \%$ catch increase. The revised 2015 assessment is now much more consistent with the corresponding 2014 and 2016 assessments. It has also transpired that the updated advice is more in line with the available fishing opportunities for the year: as of mid-November 2016, the quota uptake for the year was still under $50 \%$, indicating that the high quota advised for 2016 was not supported by the actual stock.

Mean-standardising SSB and recruitment estimates (using a common year-range for the mean) and generating TSA and SAM estimates of $Z$ by adding $F$ and $M$ enables the comparison between TSA, SAM and SURBAR shown in Figure 13.3.10. SSB and recruitment estimates are very similar from the three models, although it is noticeable that the SURBAR estimates for large year-classes in particular tend to be higher, and the swings between high and low SURBAR SSB estimates are more pronounced than for TSA and SAM: the final year SSB estimate from SAM is very similar to that from TSA. The mean Z time-series from SURBAR is consistent with that from TSA, while the SAM mean $Z$ estimates tend to be smoother, but the overall trends are not significantly different: again, we note that the final year mean $Z$ estimate from SAM is lower than that from TSA. Overall, the SAM and SURBAR assessments concur with and support the final TSA assessment, with some relatively minor variations.

### 13.3.5 Final assessment

Table 13.3.1 gives the final TSA assessment settings, while Table 13.3.2 gives the corresponding parameter estimates from the completed run. A full description of the TSA method and the purposes of each parameter are given in the Stock Annex, and the ICES-WKHAD (2014) report. Note that, for assessment purposes, total catch is divided into human consumption landings (referred to as "landings") and a composite of discards and industrial bycatch (referred to as "discards" or "discards+bycatch"), as the selectivity characteristics of these latter components are similar.
The stock summary is given in Figure 13.3.11, with the stock-recruit plot in Figure 13.3.12 and the recruitment time-series in Figure 13.3.13. The latter plot shows that the underlying mean level of recruitment has declined from the early seventies until today, and recruitment remains low in general. Furthermore, the size of sporadic, larger year classes has diminished since the large 1999 year-class. Figure 13.3.14 summarises the observed and fitted discard+bycatch proportions by age, from which the decline in discard+bycatch rates across ages 2 to 4 in recent years can be seen.

Standardised prediction errors are given in Figures 13.3.15 (landings), 13.3.16 (discard+bycatch), 13.3.17 (the IBTS Q1 survey) and 13.3.18 (the IBTS Q3 survey). These are the principal diagnostic tools for fitting time-series Kalman filter models like TSA, and indicate the discrepancy between the model prediction and observation as the model steps through the data from the start to the end. They are a useful guide to suggest observations which might need to be downweighted, but as TSA also includes a backwards smoothing step they cannot be considered to be residuals in the usual sense.

The time-series of observed and fitted values for total catch (Figure 13.3.19), the IBTS Q1 survey (Figure 13.3.20) and the IBTS Q3 survey (Figure 13.3.21) are more interpretable in that context. The estimate of total catch at age-0 prior to 1991 is based on quite
noisy discard+bycatch data where they are available, or on model inference where they are not (1973-1977), so for the earlier period model fits are not necessarily very close to observations. The other notable feature is that total catch tends to be overestimated for larger year-classes, whereas survey indices tend to be slightly underestimated for these year-classes: the TSA model fit is a compromise between the two.
Figure 13.3.22 summarises the results of TSA retrospective analyses for Northern Shelf haddock. As discussed in detail in the IBP report (ICES-IBPHaddock 2016), these show considerably less bias than the equivalent plots from the May 2016 assessment, due principally to a different treatment of recent larger year-classes.
Fishing mortality estimates for the final TSA assessment are presented in Table 13.3.3, the stock numbers in Table 13.3.4, and the assessment summary in Table 13.3.5.

### 13.4 Historical Stock Trends

The historical stock and fishery trends are presented in Figure 13.3.11.
Landings yields have stabilised since 2000, partly due (until 2014) to the limitation of inter-annual TAC variation to $\pm 15 \%$ in the EU-Norway management plan for the North Sea. Discards have fluctuated in the same period due to the appearance and subsequent growth of the 1999, 2005 and 2009 year-classes, while industrial bycatch (IBC) is now at a very low level for haddock (see also Figure 13.2.3).

Estimated fishing mortality for 2008 to 2015 appears to fluctuate between 0.2 and 0.4 and remains above the new value of F (msy) of 0.19 (see Section 13.7) in 2015. Fluctuations around the previous target-F rate (0.3) of the management plan are an expected consequence of the lag between data collection and management action, and should not be taken to indicate that the plan did not work. The 2006-2008 and 2010-2013 yearclasses are estimated to have been very weak, and the fishery has been sustained in recent years by the 2005 and 2009 year-classes. The 2014 year-class is modest in size compared to the previous sporadic larger year classes and is below the long-term average for recruitment. Therefore, it is expected to make a smaller contribution to the stock compared to other "larger" year classes over the next few years.

### 13.5 Recruitment estimates

Following the Stock Annex, recruits in the intermediate year ( $I Y=2016$ ) and in the quota year $(\mathrm{IY}+1=2017)$ are based on the TSA estimate of forecasted recruits at age 0 in the intermediate year, as this ensures consistency between assessment and forecast. At the time of the final assessment (November 2016), the results of the IBTS Q3 survey were available, and these were included in the TSA run.

The following table summarises the recruitment, age 1 and age 2 assumptions for the short term forecast.

|  | Year class | Age in 2016 | TSA estimate <br> (millions) | TSA forecast <br> (millions) |
| :--- | :--- | :--- | :--- | :--- |
| 2014 | 2 | 618 |  |  |
| 2015 | 1 | 541 | 3279 |  |
| 2016 | 0 |  | 3279 |  |
| 2017 | Age 0 in 2017 | 3279 |  |  |
| 2018 | Age 0 in 2018 |  |  |  |

### 13.6 Short-term forecasts

## Weights-at-age

Mean weights-at-age are forecast using the method proposed by Jaworski (2011) and discussed by ICES-WKHAD (2014). The method is also summarised in the Stock Annex, and involves fitting straight lines to cohort-based weight estimates and extrapolating forward in time.

The outcomes are summarized in Figures 13.6 .1 (total catch), 13.6 .2 (landings) and 13.6.3 (discards). There is insufficient data to allow for cohort-based modeling of weights-at-age in the industrial bycatch component, so simple three-year (2013-2015) means by age are used for all forecast years.

## Fishing mortality

ICES-WKHAD (2014) concluded that fishing mortality estimates for the intermediate year should be taken to be the same as the final year, considering that $F$ is smoothed within the TSA model. When this approach results in landings that overshoot the TAC, a TAC constraint should be considered: however, this is not the case for the 2016 intermediate year as the quota will not be fully utilised.
Given the choice of fishing-mortality rates discussed above, partial fishing mortality values were obtained for each catch component (human consumption, discards and bycatch) by using the relative contribution (averaged over 2013-2015) of each component to the total catch.

## Splitting catch forecasts between management units

The haddock assessment presented in this section is for the combined Northern Shelf stock, following the conclusion from ICES-WKHAD (2014) that this was biologically appropriate. However, catch advice is still required for the extant management units. ICES-WKHAD (2014) proposed a survey-based method for splitting forecast catch into sub-units on the basis of a time-smoothed survey-based estimate of the proportion of the fishable stock in each area in each year. This is summarised in the Stock Annex.

However, the survey-based proportions were not accepted by ACOM (in June 2014) as the basis for advice, due to concerns over the comparability of survey catchability between the three management areas covered by the assessment area. As a consequence, the catch forecasts provided in Table 13.6.2 are provided for the full stock area only (Subarea 4 and Divisions 3.a and 6.a).

## Forecast results

The inputs to the short-term forecast (conducted using the MFDP program) are presented in Table 13.6.1. Results for the short-term forecasts are presented in Table 13.6.2.

Assuming TSA-smoothed F in both 2016 and 2017, SSB is expected to decrease to 119 477 t in 2016, before rising in 2017 to 235113 t . In this case, human consumption yield in 2016 would will be 62993 t with associated discards of 15985 t .

Several alternative options have been highlighted in Table 13.6.2. Among these are a forecast with total fishing mortality fixed to the level specified in the previous EU-Norway North Sea management plan ( $\mathrm{F}=0.3$ ), a revised estimate of $\mathrm{F}_{\text {msy }}=0.19$ (see Section 13.7), and forecasts using a range of multipliers of $\mathrm{F}_{\mathrm{sq}}$ as the basis. Under the assumption of $\mathrm{F}_{\text {msy, }}$, the 2017 total catch is forecast to be 39461 t , which corresponds (if 2015 discard rates remain unchanged) to a wanted-catch yield of 33385 t and unwanted catch of 6071 t . This exploitation is forecast to lead in turn to SSB in 2018 of 205595 t , a decrease of $13 \%$ on the 2017 forecast.

Table 13.6.2 includes an additional column, summarizing the percentage change between the wanted catch in 2015 and the forecast total catch in 2017. The WG proposes that this may be a more indicative comparison, as it circumvents the problem of the overly optimistic and incorrect forecast used as the basis for the 2016 advice issued in June 2015, while accounting for the inclusion in 2016 of haddock in the EU Landing Obligation regulation.

### 13.7 Medium-term forecasts

No specific medium-term forecasts have been carried out for this stock. Management simulations over the medium-term period were performed for North Sea haddock (Needle 2008a, b) and West of Scotland haddock (Needle 2010), as discussed briefly in Section 13.1.4 above.

### 13.8 Biological reference points

Following the estimation of revised FMSY reference points at the 2014 WKMSYREF3 meeting, WGNSSK conducted further analysis using the EqSIM software to check that the estimated points remained valid following the update assessment. These analyses were repeated by the IBP following the modifications made to the assessment (ICESIBPHaddock 2016). Figure 13.8.1 summarises the output from this analysis, which indicates that an appropriate value of $\mathrm{Fmsy}^{\text {m for Northern Shelf haddock is now } 0.19 \text {. This }}$ is a reduction from the value set at WKMSYREF3 (0.37): the key difference in the estimates is that the calculation is based on the recruitment time-series from 2000-2015, rather than the full 1972-2015 time series. WGNSSK proposes that the former period is more appropriate, as recruitment does appear to be declining (see Figure 13.3.11) and it would be unwise to assume that a very large recruitment is likely in the near future.

Using the ICES guidelines for sporadic spawners, Blim was revised to 94 kt (the estimated SSB for 1979, the smallest stock size to produce a good recruitment), and $\mathrm{B}_{\mathrm{pa}}$ was revised to $1.4 \times$ Blim = 132 kt (which was also used as the MSY Btrigger value). An EqSim run with no advice error or rule generated $\mathrm{F}_{\lim }=\mathrm{F}_{\mathrm{p} 50}=0.38$, and $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\lim } / 1.4=0.27$. A second EqSim run with advice error but no advice rule produced an estimate of $\mathrm{F}_{\text {MSY }}=$ 0.24 with the range of 0.18 to 0.30 (Figure 13.8.1, top plot). However, an EqSim run with advice error and rule showed that $\mathrm{F}_{\mathrm{p} 05}=0.19<\mathrm{F}_{\text {MSY }}$ (Figure 13.8.1, bottom plot) so both FmSY $_{\text {and }}$ ane upper limit of the $\mathrm{F}_{\text {MSY }}$ range were constrained resulting in an $\mathrm{F}_{\text {MSY }}$ estimate of 0.19 and associated range of $0.18-0.19$.

The reference points in full from this analysis are given below:

|  | VARIABLE |  | WKHAD (2014) |
| :--- | :--- | :--- | :--- |
| $B_{\lim }$ | 63 kt | IBPHADDOCK (2016) |  |
| $\mathrm{B}_{\mathrm{pa}}$ | 88 kt | 94 kt |  |
| $\mathrm{F}_{\mathrm{lim}}$ | $\mathrm{n} / \mathrm{a}$ | 132 kt |  |
| $\mathrm{F}_{\mathrm{pa}}$ | $\mathrm{n} / \mathrm{a}$ | 0.38 |  |
| $\mathrm{~F}_{\mathrm{MSY}}$ | 0.37 | 0.27 |  |

### 13.9 Quality of the assessment

Survey data are consistent both within and between surveys, and the catch data are internally consistent. Trends in mortality from catch data and survey indices are similar. Retrospective bias in the TSA model has been significantly reduced in the current implementation, and a previous coding error has been identified and removed.

### 13.10Status of the Stock

Fishing mortality is now estimated to have remained at a relatively low level in 2015 and is now fluctuating around the historical minimum, although this remains above the most recent estimate of $\mathrm{F}_{\text {msy }}(0.19)$. Discard rates have increased slightly above the historical minimum observed in 2013, but remain low. The 2010-2013 year-classes were estimated to be weak, following the relatively strong 2009 year-class, but the 2014 yearclass is slightly larger than the recent average. Recruitment since the very large 1999 year-class has generally been low, compared with the historical time series. Spawning stock biomass is predicted to decrease during 2016 to below $\mathrm{B}_{\mathrm{pa}}(132 \mathrm{kt})$ before increasing in 2017 as the 2014 year-class matures.

### 13.11 Management Considerations

The previous EU-Norway management plan for North Sea haddock, and the EU management plan for Division 6.a haddock, are not appropriate for the Northern Shelf stock, as they relate to only a part of the full stock area. Discussions are ongoing between the EU and Norway which may establish a new management strategy on the basis of the Northern Shelf stock. However, even if agreed this will require evaluation, and in the meantime the principal basis for management of this haddock stock is the MSY approach. The survey-based proposal for splitting catch advice into management subunits, which was proposed by WGNSSK in 2014, has not been agreed by ACOM, and the split of quota into management units remains based on historical landings. It is unlikely, therefore, to follow any future changes in stock distribution across the Northern Shelf.

Considering the Northern Shelf as a whole, fishing mortality declined significantly in the early 2000s and has fluctuated around a relatively low level since. However, the current estimate remains above the proposed new value of $\mathrm{F}_{\text {msy }}$. Spawning stock biomass is estimated to have reached a historical peak in 2002 with the growth of the large 1999 year-class, but declined again rapidly and is now driven strongly by occasional moderate year-classes. The most recent of these occurred in 2005, 2009 and 2014: other recent cohorts have been very weak. SSB is likely to decline further in the future, even with low fishing mortality, until the maturation of the 2014 year-class in 2017. However, the impact on SSB of the 2014 year class is expected to be less than previous moderate year classes.

Keeping fishing mortality close to the target MSY level would be preferable to encourage the sustainable exploitation of the 2009 and 2014 year-classes. Estimated discard rates are now low, which may be due partly to the lack of small fish in the population, and partly due to an increased awareness of discard problems following public campaigns and (particularly) the installation of CCTV monitoring cameras on a number of vessels. However, discard rates do remain high in certain small-mesh fisheries (such as the TR2 Nephrops fleets in Division 6.a). Further improvements to gear selectivity measures, allowing for the release of small fish, would be highly beneficial not only for the haddock stock, but also for the survival of juveniles of other species that occur in mixed fisheries along with haddock. Similar considerations also apply to spatial management approaches (such as real-time closures), and other measures intended to reduce unwanted bycatch and discarding of various species (such as the Scottish Conservation Credits scheme; see Section 13.1.4). Haddock is included in the EU Landings Obligation regulation from 2016, though the impacts on fishing and on the stock are as yet unknown.

Haddock is a specific target for some fleets, but is also caught as part of a mixed fishery catching cod, whiting and Nephrops. It is important to consider both the species-specific assessments of these species for effective management, as well as the latest developments in the mixed fisheries approach. This is not straightforward when stocks are managed via a series of single-species, single-area management plans that do not incorporate mixed-stocks considerations. However, a reduction in effort on one stock may lead to a reduction or an increase in effort on another and the implications of any change need to be considered carefully.

### 13.12 Assessment frequency

Regarding the Northern Shelf haddock assessment, the following summarises the WGNSSK responses to each of the criteria:-

- Stocks are considered candidates for biennial assessment if the advice for the stock has been 0 -catch or equivalent for the latest three advice years.
- This does not apply for haddock.

Stocks are considered candidates for biennial assessment if the following criteria are fulfilled simultaneously.

- Life span (i.e. maximum normal age) of the species is larger than 5 years.
- This applies to haddock.
- The stock status in relation to the reference points is according to the MSY criteria F (latest assessment year) $<=1.1 \times$ Fmsy OR if Fmsy range has been defined: $F$ (latest assessment year) is <= Fupper (upper bound in F range) AND SSB(start of intermediate year) >= MSY Btrigger
- This does not apply to haddock.
- The average contribution to the catch in numbers of the recruiting year class in latest 5 years is less than $25 \%$ of the total catch in numbers. Should be calculated as the average over the latest five years of the catch in numbers of first age divided by the total catch in number by year.
- The first age in the assessment of haddock is zero. Applying the method given here, $2 \%$ of the catch is at age zero. Using age- 1 instead (which would be the recruiting age for most comparable stocks) gives $3 \%$. So the criterion applies to haddock as given.
- The retrospective pattern, based on a seven years peel of Mohn's Rho index, shows that F is consistently underestimated by more than $20 \%$. The formula to
be used in the calculations is: $\rho=\frac{1}{7} \sum_{u=Y-7}^{Y-1}\left(1-\frac{F_{u, u}}{F_{u, Y}}\right)$. The result should be $<$ 0.20 , where $F_{-}(u, u)$ is $F$ in year $u$ estimated from an assessment that ends in year $u$, and $F_{-}(u, Y)$ is the $F$ in year $u$ estimated from the most recent assessment (which ends in year Y )
- Mohn's rho for haddock is 0.22 , so this criterion does not apply.

The stability table is difficult to complete for this stock, because the stock definition changed in 2014 and the predicted catch from original component stocks is not directly comparable. In addition, neither the 2011 nor the 2012 advice included a catch prediction for 2014 - such a prediction was not made until the 2013 advice. A further complication for haddock is that the forecast must still be run using the MFDP program, because the corresponding FLR function does not yet allow for a third catch component (industrial bycatch, in this case). This should be possible within FLR, but the required development work has not yet been completed and MFDP is the only option in the meantime. The problem for this exercise is that MFDP can only carry out a standard one-year ahead forecast, rather than the two-year ahead forecast required for the frequency analysis.

Therefore, Northern Shelf haddock does not pass all the given criteria. In 2015, the stock did pass all the criteria, but WGNSSK argued that it still may not be a good candidate for less frequent assessment in any case. The reason is that stock dynamics are driven very strongly by the occasional (and completely unpredictable) appearance of large year-classes, and an assessment schedule that was unable to respond sufficiently quickly to these recruitment events would rapidly lead to a serious disjunction between the stock abundance and the available quota. In the context of the EU Landings Obligation, this would be particularly problematic. On the other hand, it generally takes two years for the recruits observed at age 0 in the IBTS Q3 survey to fully recruit to the human consumption fishery, so a two-year quota may be sufficient to account for large incoming year-classes. It is hard to be certain what the outcome would be, however, without more comprehensive risk analyses.

This leads to the more general point. One further opinion expressed during the WGNSSK discussion on this issue was that relatively simple tests would generally be insufficient to determine the risk of unwanted outcomes, should the frequency of assessments for a particular stock be reduced. Such an exercise would require a simulation analysis of the type used to evaluate management plans and strategies. An approach of this kind would take considerable time that would not be available during the WG meeting itself, and would thus require the implementation of a directed Expert Group or coordinated intersessional work. Several members of WGNSSK have tried to set up such a Group within ICES in recent years to no avail, and the difficulty of instigating this work should not be underestimated. There remains a real concern that the simple application of the criteria could lead rapidly to very undesirable outcomes which cannot be predicted without a more robust risk analysis.

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Table 13.2.1. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Nominal landings ( 000 t) during 2006-2015, as officially reported to, and estimated by, ICES, along with WG estimates of catch components, and corresponding TACs. Landings estimates for 2015 are preliminary. Quota uptake estimates are also given, calculated as the WG estimates of landings divided by available quota. Note that the United Kingdom did not provide official landings for 2012.

| Division 3.A |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Country | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| DE | 186 | 206 | 87 | 105 | 65 | 102 | 120 | 90 | 114 | 103 |
| DK | 1001 | 1054 | 1052 | 1263 | 1139 | 1661 | 1916 | 1456 | 1763 | 1057 |
| NL | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 5 | 6 | 4 |
| NO | 113 | 152 | 170 | 121 | 81 | 125 | 239 | 223 | 81 | 63 |
| PT | 30 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SE | 246 | 278 | 276 | 166 | 126 | 198 | 210 | 217 | 219 | 202 |
| UK | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 |
| Subarea 4 |  |  |  |  |  |  |  |  |  |  |
| Country | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| BE | 106 | 178 | 112 | 108 | 78 | 106 | 78 | 78 | 98 | 45 |
| DE | 726 | 727 | 393 | 657 | 634 | 575 | 548 | 677 | 677 | 599 |
| DK | 759 | 645 | 501 | 552 | 725 | 697 | 947 | 1283 | 1079 | 1426 |
| ES | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| FO | 4 | 0 | 3 | 32 | 5 | 0 | 0 | 0 | 0 | 0 |
| FR | 444 | 498 | 448 | 135 | 276 | 320 | 175 | 177 | 209 | 101 |
| GL | 5 | 8 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 |
| IE | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| IS | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| NL | 33 | 55 | 29 | 24 | 41 | 71 | 191 | 172 | 99 | 43 |
| NO | 1798 | 1706 | 1482 | 1278 | 1126 | 1195 | 1069 | 1661 | 2705 | 2004 |
| PL | 8 | 8 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| PT | 76 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SE | 100 | 130 | 83 | 141 | 90 | 128 | 103 | 113 | 154 | 135 |
| UK | 32390 | 26717 | 27365 | 28393 | 24983 | 23343 | 0 | 32993 | 29758 | 25852 |
| DIvision 6.A |  |  |  |  |  |  |  |  |  |  |
| Country | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| DE | 7 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| ES | 44 | 5 | 10 | 21 | 28 | 36 | 15 | 0 | 19 | 9 |
| FO | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| FR | 291 | 211 | 151 | 136 | 89 | 73 | 32 | 51 | 67 | 41 |
| IE | 526 | 759 | 879 | 297 | 396 | 290 | 845 | 746 | 653 | 768 |
| NO | 17 | 16 | 28 | 18 | 9 | 4 | 0 | 6 | 15 | 7 |
| UK | 4947 | 2780 | 1776 | 2380 | 2415 | 1364 | 0 | 3878 | 3230 | 3051 |
| Northern Shelf |  |  |  |  |  |  |  |  |  |  |
|  | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| Official landings | 43858 | 36172 | 34862 | 35831 | 32308 | 30288 | 6488 | 43830 | 40945 | 35520 |
| ICES landings | 43334 | 34672 | 33058 | 35590 | 31940 | 36570 | 38162 | 43681 | 41143 | 35316 |
| ICES discards | 23094 | 32651 | 14503 | 12326 | 13071 | 13067 | 5032 | 3038 | 5090 | 6255 |
| ICES IBC | 535 | 48 | 199 | 52 | 431 | 24 | 1 | 54 | 65 | 21 |
| ICES total catch | 66962 | 67371 | 47759 | 47968 | 45442 | 49661 | 43195 | 46772 | 46295 | 41571 |
| TAC 4 | 51850 | 54640 | 46444 | 42110 | 35794 | 34057 | 39000 | 45041 | 38284 | 40711 |
| TAC 3.a | 3189 | 3360 | 2856 | 2590 | 2201 | 2100 | 2095 | 2770 | 2355 | 2504 |
| TAC 6.a | 7810 | 7200 | 6120 | 3520 | 2670 | 2005 | 6015 | 4211 | 3988 | 4536 |
| Total TAC | 62849 | 65200 | 55420 | 48220 | 40665 | 38162 | 47110 | 52022 | 44627 | 47751 |
| ICES quota uptake | 69\% | 53\% | 60\% | 74\% | 79\% | 96\% | 81\% | 84\% | 92\% | 74\% |

Table 13.2.2. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Proportion of sampling strata for discards imported into InterCatch and proportion of discards raised from averaged discard rates.

| Catch category | Raised or imported | Weight (tonnes) | Proportion |
| :--- | :--- | :--- | :--- | :--- |
| DISCARDS | RAISED | 610622 | 10 |
| DISCARDS | IMPORTED | 5583761 | 90 |
| LANDINGS | IMPORTED | 35490794 | 100 |

Table 13.2.3 Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Proportion of age distributions for landings and discards either imported or raised in InterCatch and either sampled or estimated.

| Catch category | Raised or <br> imported | Sampled or <br> estimated | Weight (tonnes) | Proportion |
| :--- | :--- | :--- | :--- | :--- |
| LANDINGS | IMPORTED | SAMPLED | 31776693 | 90 |
| LANDINGS | IMPORTED | ESTIMATED | 3539434 | 10 |
| DISCARDS | IMPORTED | SAMPLED | 5508464 | 88 |
| DISCARDS | RAISED | ESTIMATED | 616607 | 10 |
| DISCARDS | IMPORTED | ESTIMATED | 130143 | 2 |

Table 13.2.4 Haddock in Subarea 4, Division 3.a. 2 and Division 6.a. Proportion by area of distributions for landings and discards either imported or raised in InterCatch and either sampled or estimated.

| Catch category | Raised or imported | Sampled or estimated | Area | Weight (tonnes) | Proportion |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | Imported | Sampled | IIIaN | 1041917 | 73 |
| Landings | Imported | Estimated | IIIaN | 380466.8 | 27 |
| Discards | Raised | Estimated | IIIaN | 58622.38 | 41 |
| Discards | Imported | Sampled | IIIaN | 50341.03 | 35 |
| Discards | Imported | Estimated | IIIaN | 35591.27 | 2 |
| Landings | Imported | Sampled | IV | 25595184 | 90 |
| Landings | Imported | Estimated | IV | 2877496 | 10 |
| Discards | Imported | Sampled | IV | 4128573 | 90 |
| Discards | Raised | Estimated | IV | 452611 | 10 |
| Discards | Imported | Estimated | IV | 15566.43 | 0 |
| Landings | Imported | Sampled | IVa | 1180255 | 99 |
| Landings | Imported | Estimated | IVa | 15471.65 | 1 |
| Discards | Raised | Estimated | IVa | 34601.67 | 58 |
| Discards | Imported | Sampled | IVa | 18180.07 | 30 |
| Discards | Imported | Estimated | IVa | 6959.729 | 12 |
| Landings | Imported | Estimated | IVb | 195995.6 | 71 |
| Landings | Imported | Sampled | IVb | 80037.57 | 29 |
| Discards | Raised | Estimated | IVb | 13529.61 | 39 |
| Discards | Imported | Estimated | IVb | 11143.41 | 32 |
| Discards | Imported | Sampled | IVb | 9786.963 | 28 |
| Discards | Imported | Estimated | IVc | 0 | NA |
| Landings | Imported | Sampled | VIa | 3879299 | 98 |
| Landings | Imported | Estimated | VIa | 70003.73 | 2 |
| Discards | Imported | Sampled | VIa | 1301582 | 92 |
| Discards | Raised | Estimated | VIa | 57241.97 | 4 |
| Discards | Imported | Estimated | VIa | 60882.39 | 4 |

Table 13.2.5. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Working Group estimates of catch components by weight ( 000 tonnes). *Note that Subarea 4 and Division 3.a. 20 data are collated together in 2013, and are listed here only in the Subarea 4 section.

| Subarea 4 |  |  |  | DIVISION 3.A |  |  |  | DIVISION 6.A |  |  | Combined |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings | Discards | IBC | Total | Landings | Discards | Total | Landings | Discards | Total | Landings | Discards | IBC |  |
| 1965 | 161.7 | 62.3 | 74.6 | 298.6 | 0.7 |  | 0.7 | 32.5 | 3.4 | 35.9 | 194.9 | 65.7 | 74.6 | 335.2 |
| 1966 | 225.6 | 73.5 | 46.7 | 345.8 | 0.6 |  | 0.6 | 29.9 | 0.7 | 30.6 | 256.1 | 74.2 | 46.7 | 377.0 |
| 1967 | 147.4 | 78.2 | 20.7 | 246.3 | 0.4 |  | 0.4 | 20.3 | 7.4 | 27.7 | 168.1 | 85.6 | 20.7 | 274.4 |
| 1968 | 105.4 | 161.8 | 34.2 | 301.4 | 0.4 |  | 0.4 | 20.5 | 25.3 | 45.8 | 126.3 | 187.1 | 34.2 | 347.6 |
| 1969 | 331.1 | 260.1 | 338.4 | 929.5 | 0.5 |  | 0.5 | 26.3 | 25.2 | 51.5 | 357.9 | 285.3 | 338.4 | 981.6 |
| 1970 | 524.1 | 101.3 | 179.7 | 805.1 | 0.7 |  | 0.7 | 34.1 | 6.2 | 40.3 | 558.9 | 107.5 | 179.7 | 846.1 |
| 1971 | 235.5 | 177.8 | 31.5 | 444.8 | 2 |  | 2 | 46.3 | 12.2 | 58.5 | 283.8 | 190.0 | 31.5 | 505.3 |
| 1972 | 193 | 128 | 29.6 | 350.5 | 2.6 |  | 2.6 | 41.1 | 16.4 | 57.5 | 236.7 | 144.4 | 29.6 | 410.7 |
| 1973 | 178.7 | 114.7 | 11.3 | 304.7 | 2.9 |  | 2.9 | 28.8 | 11.4 | 40.2 | 210.4 | 126.1 | 11.3 | 347.8 |
| 1974 | 149.6 | 166.4 | 47.5 | 363.5 | 3.5 |  | 3.5 | 18.0 | 15.4 | 33.3 | 171.1 | 181.8 | 47.5 | 400.3 |
| 1975 | 146.6 | 260.4 | 41.5 | 448.4 | 4.8 |  | 4.8 | 13.7 | 33.0 | 46.6 | 165.1 | 293.4 | 41.5 | 499.9 |
| 1976 | 165.7 | 154.5 | 48.2 | 368.3 | 7 |  | 7 | 18.8 | 15.3 | 34.1 | 191.5 | 169.8 | 48.2 | 409.5 |
| 1977 | 137.3 | 44.4 | 35 | 216.7 | 7.8 |  | 7.8 | 19.3 | 4.4 | 23.7 | 164.4 | 48.8 | 35 | 248.2 |
| 1978 | 85.8 | 76.8 | 10.9 | 173.5 | 5.9 |  | 5.9 | 17.2 | 1.1 | 18.3 | 108.9 | 77.9 | 10.9 | 197.7 |
| 1979 | 83.1 | 41.7 | 16.2 | 141 | 4 |  | 4 | 14.8 | 6.5 | 21.3 | 101.9 | 48.2 | 16.2 | 166.3 |
| 1980 | 98.6 | 94.6 | 22.5 | 215.7 | 6.4 |  | 6.4 | 12.8 | 4.8 | 17.5 | 117.8 | 99.4 | 22.5 | 239.6 |
| 1981 | 129.6 | 60.1 | 17 | 206.7 | 6.6 |  | 6.6 | 18.2 | 7.1 | 25.3 | 154.4 | 67.2 | 17 | 238.6 |
| 1982 | 165.8 | 40.6 | 19.4 | 225.8 | 7.5 |  | 7.5 | 29.6 | 7.7 | 37.3 | 202.9 | 48.3 | 19.4 | 270.6 |
| 1983 | 159.3 | 66 | 12.9 | 238.2 | 6 |  | 6 | 29.4 | 3.4 | 32.8 | 194.7 | 69.4 | 12.9 | 277.0 |
| 1984 | 128.2 | 75.3 | 10.1 | 213.6 | 5.4 |  | 5.4 | 30.0 | 8.1 | 38.1 | 163.6 | 83.4 | 10.1 | 257.1 |
| 1985 | 158.6 | 85.2 | 6 | 249.8 | 5.6 |  | 5.6 | 24.4 | 10.7 | 35.1 | 188.6 | 95.9 | 6 | 290.5 |
| 1986 | 165.6 | 52.2 | 2.6 | 220.4 | 2.7 |  | 2.7 | 19.6 | 5.2 | 24.7 | 187.9 | 57.4 | 2.6 | 247.8 |
| 1987 | 108 | 59.1 | 4.4 | 171.6 | 2.3 |  | 2.3 | 27.0 | 11.1 | 38.1 | 137.3 | 70.2 | 4.4 | 211.9 |
| 1988 | 105.1 | 62.1 | 4 | 171.2 | 1.9 |  | 1.9 | 21.1 | 5.0 | 26.1 | 128.1 | 67.1 | 4 | 199.2 |
| 1989 | 76.2 | 25.7 | 2.4 | 104.2 | 2.3 |  | 2.3 | 16.7 | 2.5 | 19.2 | 95.2 | 28.2 | 2.4 | 125.8 |
| 1990 | 51.5 | 32.6 | 2.6 | 86.6 | 2.3 |  | 2.3 | 10.1 | 0.8 | 11.0 | 63.9 | 33.4 | 2.6 | 100.0 |
| 1991 | 44.7 | 40.2 | 5.4 | 90.2 | 3.1 |  | 3.1 | 10.6 | 4.8 | 15.3 | 58.4 | 45.0 | 5.4 | 108.7 |
| 1992 | 70.2 | 47.9 | 10.9 | 129.1 | 2.6 |  | 2.6 | 11.3 | 3.5 | 14.9 | 84.1 | 51.4 | 10.9 | 146.5 |
| 1993 | 79.6 | 79.6 | 10.8 | 169.9 | 2.6 |  | 2.6 | 19.1 | 7.0 | 26.1 | 101.3 | 86.6 | 10.8 | 198.7 |
| 1994 | 80.9 | 65.4 | 3.6 | 149.8 | 1.2 |  | 1.2 | 14.2 | 5.0 | 19.2 | 96.3 | 70.4 | 3.6 | 170.3 |
| 1995 | 75.3 | 57.4 | 7.7 | 140.4 | 2.2 |  | 2.2 | 12.4 | 7.7 | 20.0 | 89.9 | 65.1 | 7.7 | 162.6 |
| 1996 | 76 | 72.5 | 5 | 153.5 | 3.1 |  | 3.1 | 13.5 | 7.8 | 21.3 | 92.6 | 80.3 | 5 | 177.9 |
| 1997 | 79.1 | 52.1 | 6.7 | 137.9 | 3.4 |  | 3.4 | 12.9 | 7.5 | 20.4 | 95.4 | 59.6 | 6.7 | 161.7 |
| 1998 | 77.3 | 45.2 | 5.1 | 127.6 | 3.8 |  | 3.8 | 14.4 | 7.0 | 21.4 | 95.5 | 52.2 | 5.1 | 152.8 |
| 1999 | 64.2 | 42.6 | 3.8 | 110.7 | 1.4 |  | 1.4 | 10.4 | 3.9 | 14.3 | 76.0 | 46.5 | 3.8 | 126.3 |
| 2000 | 46.1 | 48.8 | 8.1 | 103 | 1.5 |  | 1.5 | 7.0 | 6.3 | 13.2 | 54.6 | 55.1 | 8.1 | 117.7 |
| 2001 | 39 | 118.3 | 7.9 | 165.2 | 1.9 |  | 1.9 | 6.7 | 8.5 | 15.2 | 47.6 | 126.8 | 7.9 | 182.3 |
| 2002 | 54.2 | 45.9 | 3.7 | 103.8 | 4.1 |  | 4.1 | 7.1 | 9.4 | 16.5 | 65.4 | 55.3 | 3.7 | 124.4 |
| 2003 | 40.1 | 23.5 | 1.1 | 64.8 | 1.8 | 0.2 | 2 | 5.3 | 4.5 | 9.8 | 47.2 | 28.2 | 1.1 | 76.5 |
| 2004 | 47.3 | 15.4 | 0.6 | 63.2 | 1.4 | 0.1 | 1.6 | 3.2 | 4.5 | 7.7 | 51.9 | 20.0 | 0.6 | 72.5 |
| 2005 | 47.6 | 8.4 | 0.2 | 56.2 | 0.8 | 0.2 | 1 | 3.1 | 3.8 | 6.9 | 51.5 | 12.4 | 0.2 | 64.1 |
| 2006 | 36.1 | 16.9 | 0.5 | 53.6 | 1.5 | 1 | 2.5 | 5.7 | 5.2 | 10.9 | 43.3 | 23.1 | 0.5 | 66.9 |
| 2007 | 29.4 | 27.8 | 0 | 57.3 | 1.5 | 0.8 | 2.3 | 3.7 | 4.0 | 7.8 | 34.6 | 32.6 | 0 | 67.3 |
| 2008 | 28.9 | 12.5 | 0.2 | 41.6 | 1.4 | 0.6 | 2 | 2.8 | 1.3 | 4.1 | 33.1 | 14.4 | 0.2 | 47.7 |
| 2009 | 31.3 | 10 | 0.1 | 41.3 | 1.5 | 0.6 | 2.1 | 2.8 | 1.8 | 4.6 | 35.6 | 12.4 | 0.1 | 48.1 |
| 2010 | 27.8 | 9.5 | 0.4 | 37.7 | 1.3 | 0.6 | 1.9 | 2.9 | 2.9 | 5.8 | 32.0 | 13.0 | 0.4 | 45.4 |
| 2011 | 26.3 | 10.2 | 0 | 36.5 | 9.9 | 1.7 | 11.6 | 1.7 | 1.5 | 3.3 | 37.9 | 13.4 | 0 | 51.4 |
| 2012 | 30.3 | 3.7 | 1.2 | 35.0 | 2.6 | 0.7 | 3.4 | 5.1 | 0.5 | 5.6 | 38.0 | 4.9 | 1.2 | 44.1 |
| 2013* | 38.9 | 2.0 | 0.1 | 41.0 |  |  |  | 4.7 | 1.1 | 5.8 | 43.7 | 3.0 | 0.1 | 46.8 |
| 2014 | 34.9 | 4.1 | 0.1 | 39.1 | 2.3 | 0.1 | 2.4 | 4.0 | 0.8 | 4.8 | 41.1 | 5.1 | 0.1 | 46.3 |
| 2015 | 30.2 | 4.2 | 0.0 | 34.3 | 1.4 | 0.1 | 1.5 | 3.9 | 1.3 | 5.2 | 35.3 | 6.3 | 0.0 | 41.6 |

Table 13.2.6. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Numbers at age data (thousands) for total catch. Ages 0-7 and 8+ and years 1972-2015 are used in the assessment.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 650218 | 368560 | 16491 | 721514 | 36301 | 4954 | 2245 | 626 | 118 | 97 | 47 | 0 | 0 | 0 | 0 | 0 | 262 |
| 1966 | 1672925 | 1007517 | 26186 | 7536 | 459941 | 11903 | 1109 | 633 | 222 | 90 | 23 | 2 | 0 | 0 | 0 | 0 | 337 |
| 1967 | 345371 | 856339 | 108401 | 5814 | 3850 | 202830 | 2843 | 223 | 231 | 61 | 34 | 0 | 0 | 0 | 0 | 0 | 326 |
| 1968 | 11133 | 1226448 | 477603 | 22671 | 2303 | 3210 | 60034 | 1052 | 84 | 22 | 5 | 0 | 0 | 0 | 0 | 0 | 111 |
| 1969 | 75301 | 20554 | 3736629 | 313593 | 9029 | 2678 | 2894 | 23704 | 392 | 32 | 7 | 0 | 0 | 0 | 0 | 0 | 431 |
| 1970 | 941790 | 272467 | 218881 | 2003201 | 60200 | 1350 | 1285 | 401 | 6539 | 81 | 13 | 19 | 0 | 0 | 0 | 0 | 6652 |
| 1971 | 337277 | 1881729 | 74866 | 50845 | 480381 | 10916 | 589 | 201 | 167 | 1767 | 176 | 3 | 5 | 0 | 0 | 0 | 2119 |
| 1972 | 255110 | 696714 | 671965 | 43309 | 23547 | 211817 | 4067 | 241 | 53 | 27 | 475 | 11 | 0 | 0 | 0 | 0 | 566 |
| 1973 | 79461 | 412305 | 587335 | 260080 | 6450 | 5689 | 72652 | 1406 | 140 | 34 | 234 | 49 | 5 | 0 | 0 | 0 | 462 |
| 1974 | 665110 | 1283252 | 187149 | 342628 | 60523 | 1956 | 1795 | 22380 | 345 | 57 | 63 | 4 | 7 | 4 | 0 | 0 | 480 |
| 1975 | 51796 | 2276937 | 673960 | 62175 | 112242 | 17691 | 1078 | 718 | 6168 | 339 | 70 | 11 | 0 | 8 | 0 | 0 | 6596 |
| 1976 | 171400 | 192030 | 1127520 | 225532 | 11538 | 32677 | 5864 | 228 | 84 | 1863 | 64 | 3 | 5 | 0 | 0 | 0 | 2019 |
| 1977 | 119506 | 263702 | 109480 | 426291 | 45756 | 4984 | 6757 | 1608 | 163 | 40 | 460 | 8 | 0 | 1 | 0 | 0 | 672 |
| 1978 | 281785 | 223294 | 130963 | 31141 | 144703 | 11791 | 1582 | 2322 | 740 | 122 | 33 | 275 | 16 | 2 | 0 | 0 | 1188 |
| 1979 | 844410 | 261156 | 220200 | 45487 | 7978 | 38097 | 3069 | 377 | 629 | 181 | 57 | 13 | 52 | 3 | 0 | 0 | 935 |
| 1980 | 374573 | 439674 | 374310 | 80225 | 11364 | 2040 | 11143 | 827 | 143 | 168 | 96 | 34 | 9 | 7 | 1 | 0 | 457 |
| 1981 | 645352 | 116229 | 430149 | 180553 | 17044 | 2225 | 497 | 3320 | 164 | 78 | 26 | 32 | 5 | 1 | 4 | 0 | 311 |
| 1982 | 275508 | 217834 | 89989 | 390347 | 49835 | 4275 | 820 | 551 | 1072 | 60 | 28 | 8 | 2 | 2 | 0 | 0 | 1172 |
| 1983 | 513034 | 148158 | 222772 | 83199 | 166812 | 20055 | 2365 | 338 | 255 | 385 | 93 | 21 | 4 | 4 | 0 | 0 | 763 |
| 1984 | 95862 | 483045 | 139887 | 143821 | 29321 | 56077 | 6238 | 967 | 127 | 84 | 185 | 19 | 5 | 1 | 1 | 0 | 423 |
| 1985 | 127003 | 161400 | 441785 | 80605 | 41508 | 7082 | 18393 | 1929 | 296 | 56 | 29 | 144 | 9 | 0 | 0 | 1 | 535 |
| 1986 | 45703 | 137091 | 144075 | 328016 | 29497 | 10595 | 1686 | 4421 | 581 | 156 | 56 | 47 | 37 | 16 | 4 | 1 | 898 |
| 1987 | 10249 | 253236 | 259369 | 56407 | 92705 | 6214 | 3993 | 1187 | 2596 | 462 | 56 | 65 | 35 | 32 | 17 | 8 | 3271 |
| 1988 | 16679 | 33092 | 424014 | 96795 | 17161 | 27728 | 2030 | 874 | 368 | 1076 | 95 | 21 | 12 | 13 | 17 | 1 | 1603 |
| 1989 | 19587 | 51743 | 43162 | 216359 | 21015 | 4189 | 7671 | 763 | 285 | 170 | 469 | 69 | 8 | 3 | 2 | 1 | 1007 |
| 1990 | 19286 | 82571 | 78881 | 17811 | 60888 | 4373 | 1104 | 1839 | 254 | 100 | 54 | 13 | 12 | 1 | 4 | 2 | 439 |
| 1991 | 128703 | 188087 | 101425 | 24822 | 4706 | 17618 | 1388 | 684 | 1024 | 171 | 65 | 11 | 11 | 1 | 2 | 2 | 1287 |
| 1992 | 277933 | 166550 | 255051 | 43257 | 7162 | 1486 | 6376 | 611 | 337 | 401 | 149 | 22 | 6 | 2 | 0 | 0 | 918 |
| 1993 | 136841 | 302610 | 269220 | 123469 | 11822 | 1986 | 669 | 2050 | 215 | 210 | 188 | 84 | 4 | 4 | 0 | 0 | 706 |
| 1994 | 89104 | 91674 | 339428 | 106673 | 35056 | 3381 | 601 | 366 | 746 | 132 | 48 | 36 | 26 | 5 | 0 | 0 | 992 |
| 1995 | 200151 | 336460 | 119210 | 182969 | 33802 | 9237 | 898 | 161 | 155 | 151 | 21 | 8 | 6 | 2 | 1 | 0 | 345 |
| 1996 | 167032 | 46797 | 505401 | 73987 | 66245 | 11159 | 4058 | 1080 | 75 | 72 | 37 | 9 | 8 | 3 | 1 | 0 | 205 |
| 1997 | 36954 | 162449 | 107657 | 251339 | 18037 | 18288 | 2762 | 937 | 121 | 16 | 18 | 5 | 4 | 4 | 2 | 0 | 170 |
| 1998 | 21919 | 88387 | 224037 | 60861 | 128348 | 7110 | 4590 | 850 | 263 | 60 | 7 | 8 | 3 | 2 | 1 | 1 | 345 |
| 1999 | 90634 | 69455 | 119094 | 110046 | 28510 | 45221 | 2700 | 2047 | 438 | 53 | 8 | 3 | 3 | 2 | 0 | 0 | 507 |
| 2000 | 12630 | 397390 | 110381 | 61263 | 33137 | 7254 | 9935 | 765 | 367 | 53 | 13 | 2 | 1 | 1 | 0 | 0 | 438 |
| 2001 | 3518 | 95086 | 633162 | 34548 | 12078 | 5573 | 2094 | 1611 | 257 | 89 | 28 | 3 | 4 | 0 | 0 | 0 | 382 |
| 2002 | 50927 | 36063 | 99685 | 372036 | 7812 | 2801 | 1615 | 729 | 603 | 283 | 25 | 8 | 5 | 0 | 0 | 0 | 923 |
| 2003 | 7082 | 13136 | 15234 | 48729 | 127241 | 2166 | 786 | 339 | 144 | 100 | 48 | 5 | 1 | 0 | 0 | 0 | 299 |
| 2004 | 3758 | 25698 | 24627 | 8958 | 38784 | 97827 | 1010 | 248 | 82 | 42 | 37 | 12 | 1 | 0 | 0 | 0 | 174 |
| 2005 | 8779 | 17695 | 24596 | 15085 | 5446 | 27745 | 61457 | 371 | 132 | 38 | 11 | 8 | 4 | 1 | 0 | 0 | 193 |
| 2006 | 3229 | 122537 | 30995 | 20657 | 11284 | 6078 | 16415 | 32978 | 156 | 56 | 20 | 7 | 4 | 1 | 0 | 0 | 243 |
| 2007 | 2046 | 20565 | 171600 | 16796 | 8187 | 4782 | 2237 | 6876 | 7254 | 75 | 8 | 14 | 3 | 1 | 0 | 0 | 7355 |
| 2008 | 3780 | 15005 | 31864 | 75341 | 4757 | 2050 | 1516 | 566 | 1432 | 2570 | 5 | 8 | 1 | 1 | 0 | 0 | 4017 |
| 2009 | 10483 | 11042 | 15303 | 20764 | 78513 | 1860 | 845 | 567 | 239 | 276 | 569 | 6 | 2 | 0 | 0 | 0 | 1092 |
| 2010 | 2930 | 108139 | 17377 | 17834 | 11301 | 38134 | 853 | 416 | 160 | 83 | 85 | 148 | 9 | 0 | 0 | 3 | 488 |
| 2011 | 3003 | 6082 | 66355 | 17091 | 14138 | 11495 | 23124 | 677 | 282 | 95 | 17 | 5 | 60 | 0 | 0 | 0 | 459 |
| 2012 | 1319 | 3389 | 5260 | 66109 | 5388 | 3670 | 2416 | 7900 | 157 | 178 | 68 | 44 | 57 | 24 | 4 | 0 | 532 |
| 2013 | 1285 | 11998 | 4394 | 4838 | 68899 | 2269 | 1539 | 879 | 3896 | 37 | 7 | 8 | 2 | 2 | 2 | 0 | 3954 |
| 2014 | 3537 | 7504 | 19838 | 4818 | 7799 | 46760 | 1104 | 980 | 390 | 1706 | 14 | 6 | 1 | 1 | 0 | 2 | 2121 |
| 2015 | 3820 | 27637 | 15799 | 17624 | 1730 | 5166 | 22109 | 1059 | 433 | 437 | 782 | 107 | 0 | 0 | 0 | 0 | 1759 |

Table 13.2.7. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Numbers at age data (thousands) for landings. Ages 0-7 and 8+ are used in the assessment.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ | $8+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 0 | 2670 | 3908 | 396363 | 30232 | 4358 | 2126 | 620 | 118 | 97 | 47 | 0 | 0 | 0 | 0 | 0 | 262 |
| 1966 | 0 | 13034 | 6899 | 5332 | 419437 | 11113 | 1082 | 631 | 222 | 90 | 23 | 2 | 0 | 0 | 0 | 0 | 337 |
| 1967 | 0 | 55548 | 40030 | 4627 | 3607 | 198991 | 2821 | 223 | 231 | 61 | 34 | 0 | 0 | 0 | 0 | 0 | 326 |
| 1968 | 0 | 22108 | 151474 | 17130 | 2160 | 3176 | 59110 | 1051 | 84 | 22 | 5 | 0 | 0 | 0 | 0 | 0 | 111 |
| 1969 | 0 | 143 | 759680 | 175763 | 7965 | 2282 | 2760 | 23452 | 392 | 32 | 7 | 0 | 0 | 0 | 0 | 0 | 431 |
| 1970 | 0 | 2428 | 52031 | 1211535 | 53570 | 1184 | 1220 | 398 | 6539 | 81 | 13 | 19 | 0 | 0 | 0 | 0 | 6652 |
| 1971 | 0 | 35945 | 27011 | 37832 | 448352 | 10551 | 582 | 201 | 167 | 1767 | 176 | 3 | 5 | 0 | 0 | 0 | 2119 |
| 1972 | 0 | 13354 | 233966 | 35440 | 22165 | 210167 | 4054 | 241 | 53 | 27 | 475 | 11 | 0 | 0 | 0 | 0 | 566 |
| 1973 | 0 | 7277 | 211018 | 209961 | 6085 | 5459 | 72528 | 1406 | 140 | 34 | 234 | 49 | 5 | 0 | 0 | 0 | 462 |
| 1974 | 0 | 25699 | 55734 | 236624 | 53054 | 1868 | 1679 | 22156 | 345 | 57 | 63 | 4 | 7 | 4 | 0 | 0 | 480 |
| 1975 | 0 | 28773 | 211495 | 41030 | 93617 | 17406 | 1073 | 718 | 6163 | 339 | 70 | 11 | 0 | 8 | 0 | 0 | 6591 |
| 1976 | 0 | 3045 | 246027 | 155162 | 11292 | 29594 | 5846 | 228 | 84 | 1863 | 64 | 3 | 5 | 0 | 0 | 0 | 2019 |
| 1977 | 0 | 8934 | 33058 | 278741 | 42737 | 4737 | 6516 | 1608 | 163 | 40 | 460 | 8 | 0 | 1 | 0 | 0 | 672 |
| 1978 | 0 | 13913 | 55636 | 26119 | 123655 | 11479 | 1496 | 2317 | 740 | 122 | 33 | 275 | 16 | 2 | 0 | 0 | 1187 |
| 1979 | 0 | 16077 | 120456 | 38247 | 7752 | 37353 | 3052 | 377 | 629 | 181 | 57 | 13 | 52 | 3 | 0 | 0 | 935 |
| 1980 | 0 | 11487 | 154765 | 67241 | 9978 | 1985 | 11057 | 820 | 143 | 166 | 96 | 34 | 9 | 7 | 1 | 0 | 456 |
| 1981 | 0 | 1959 | 174018 | 128102 | 16447 | 2219 | 494 | 3320 | 164 | 78 | 26 | 32 | 5 | 1 | 4 | 0 | 311 |
| 1982 | 0 | 7623 | 40161 | 282492 | 45732 | 3811 | 820 | 551 | 1072 | 60 | 28 | 8 | 2 | 2 | 0 | 0 | 1172 |
| 1983 | 0 | 7669 | 114118 | 57151 | 152477 | 19147 | 2201 | 338 | 255 | 385 | 93 | 21 | 4 | 4 | 0 | 0 | 763 |
| 1984 | 0 | 22842 | 80349 | 115405 | 27331 | 52226 | 6238 | 967 | 127 | 84 | 185 | 19 | 5 | 1 | 1 | 0 | 423 |
| 1985 | 0 | 3059 | 267559 | 75242 | 40846 | 6858 | 18360 | 1929 | 296 | 56 | 29 | 144 | 9 | 0 | 0 | 1 | 535 |
| 1986 | 0 | 12735 | 67173 | 287995 | 29371 | 10587 | 1685 | 4421 | 581 | 156 | 56 | 47 | 37 | 16 | 4 | 1 | 898 |
| 1987 | 0 | 11150 | 120584 | 46970 | 89772 | 6212 | 3993 | 1187 | 2596 | 462 | 56 | 65 | 35 | 32 | 17 | 8 | 3271 |
| 1988 | 0 | 2371 | 167090 | 83798 | 16114 | 27515 | 2030 | 874 | 344 | 1076 | 95 | 21 | 12 | 13 | 17 | 1 | 1579 |
| 1989 | 0 | 5446 | 17801 | 146467 | 19506 | 4130 | 7549 | 752 | 283 | 170 | 467 | 69 | 8 | 3 | 2 | 1 | 1003 |
| 1990 | 0 | 6279 | 46366 | 15680 | 54465 | 4117 | 1054 | 1761 | 250 | 100 | 54 | 13 | 12 | 1 | 4 | 2 | 435 |
| 1991 | 0 | 21627 | 57480 | 23058 | 4646 | 17468 | 1388 | 684 | 1024 | 171 | 65 | 11 | 11 | 1 | 2 | 2 | 1287 |
| 1992 | 0 | 3544 | 128147 | 38838 | 7038 | 1483 | 6354 | 611 | 337 | 401 | 149 | 22 | 6 | 2 | 0 | 0 | 918 |
| 1993 | 0 | 3232 | 92828 | 102781 | 11570 | 1976 | 669 | 2028 | 215 | 210 | 188 | 84 | 4 | 4 | 0 | 0 | 706 |
| 1994 | 0 | 1484 | 75783 | 85391 | 32827 | 3345 | 600 | 366 | 746 | 132 | 48 | 36 | 26 | 5 | 0 | 0 | 992 |
| 1995 | 0 | 2410 | 32846 | 114437 | 31198 | 9038 | 898 | 161 | 155 | 151 | 21 | 8 | 6 | 2 | 1 | 0 | 345 |
| 1996 | 0 | 1179 | 84349 | 41653 | 55794 | 11123 | 4058 | 1080 | 75 | 72 | 37 | 9 | 8 | 3 | 1 | 0 | 205 |
| 1997 | 0 | 2292 | 26774 | 140099 | 16153 | 17846 | 2762 | 937 | 121 | 16 | 18 | 5 | 4 | 4 | 2 | 0 | 170 |
| 1998 | 0 | 2167 | 45449 | 42411 | 106125 | 6959 | 4579 | 850 | 263 | 60 | 7 | 8 | 3 | 2 | 1 | 1 | 345 |
| 1999 | 0 | 1340 | 31357 | 60351 | 26260 | 42494 | 2648 | 2047 | 438 | 53 | 8 | 3 | 3 | 2 | 0 | 0 | 507 |
| 2000 | 0 | 5508 | 32823 | 34517 | 27247 | 6927 | 9734 | 765 | 367 | 53 | 13 | 2 | 1 | 1 | 0 | 0 | 438 |
| 2001 | 0 | 855 | 75731 | 17938 | 10929 | 5321 | 2094 | 1609 | 256 | 89 | 28 | 3 | 4 | 0 | 0 | 0 | 381 |
| 2002 | 0 | 816 | 14893 | 124903 | 6330 | 2710 | 1615 | 618 | 603 | 283 | 25 | 8 | 5 | 0 | 0 | 0 | 923 |
| 2003 | 0 | 53 | 2119 | 16076 | 81868 | 2141 | 777 | 339 | 144 | 100 | 48 | 5 | 1 | 0 | 0 | 0 | 299 |
| 2004 | 0 | 495 | 3142 | 4906 | 23978 | 77262 | 996 | 239 | 82 | 42 | 37 | 12 | 1 | 0 | 0 | 0 | 174 |
| 2005 | 0 | 788 | 5777 | 8878 | 4178 | 22915 | 56760 | 370 | 131 | 38 | 11 | 8 | 4 | 1 | 0 | 0 | 192 |
| 2006 | 0 | 2129 | 10416 | 11780 | 8602 | 5209 | 14745 | 30350 | 149 | 54 | 20 | 7 | 3 | 1 | 0 | 0 | 234 |
| 2007 | 0 | 1146 | 28873 | 11204 | 7361 | 4684 | 2199 | 6773 | 7183 | 75 | 8 | 14 | 3 | 1 | 0 | 0 | 7284 |
| 2008 | 0 | 299 | 6472 | 50965 | 4461 | 1986 | 1378 | 563 | 1402 | 2566 | 5 | 8 | 1 | 1 | 0 | 0 | 3983 |
| 2009 | 0 | 486 | 4605 | 9666 | 61972 | 1775 | 793 | 521 | 239 | 276 | 566 | 6 | 2 | 0 | 0 | 0 | 1088 |
| 2010 | 0 | 1089 | 5150 | 12597 | 10176 | 35718 | 828 | 416 | 146 | 83 | 85 | 147 | 9 | 0 | 0 | 3 | 473 |
| 2011 | 0 | 224 | 16505 | 15260 | 13321 | 11383 | 22889 | 677 | 282 | 95 | 16 | 5 | 60 | 0 | 0 | 0 | 458 |
| 2012 | 0 | 261 | 3286 | 52091 | 4884 | 3660 | 2408 | 7885 | 157 | 178 | 68 | 44 | 57 | 24 | 4 | 0 | 532 |
| 2013 | 0 | 983 | 2493 | 4338 | 66123 | 2240 | 1526 | 867 | 3868 | 37 | 6 | 8 | 2 | 2 | 2 | 0 | 3924 |
| 2014 | 0 | 232 | 12630 | 3832 | 7626 | 42509 | 1100 | 965 | 382 | 1703 | 14 | 6 | 1 | 1 | 0 | 2 | 2110 |
| 2015 | 0 | 717 | 10574 | 16080 | 1636 | 5135 | 21121 | 1059 | 433 | 437 | 780 | 107 | 0 | 0 | 0 | 0 | 1758 |

Table 13.2.8. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Numbers-at-age data (thousands) for discards. Ages 0-7 and 8+ are used in the assessment.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ | $8+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 5757 | 111654 | 4897 | 141863 | 3704 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1966 | 13832 | 445648 | 12742 | 1197 | 24643 | 35 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1967 | 46372 | 408281 | 62831 | 1032 | 219 | 1576 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1968 | 67 | 741402 | 244976 | 3512 | 97 | 15 | 186 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1969 | 4475 | 5234 | 1273332 | 39179 | 432 | 16 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1970 | 68905 | 99125 | 78340 | 306391 | 2663 | 13 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 14189 | 1275394 | 37883 | 9623 | 25648 | 66 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 18446 | 444794 | 380988 | 6846 | 1236 | 1212 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 38129 | 287558 | 363916 | 50108 | 354 | 33 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 88456 | 982287 | 99148 | 59143 | 2869 | 6 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 7479 | 1653311 | 377845 | 16385 | 13423 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 6418 | 122012 | 698428 | 41183 | 200 | 137 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 16364 | 107748 | 47070 | 79922 | 664 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 1193 | 83683 | 63997 | 4214 | 19568 | 248 | 80 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 4795 | 119245 | 82074 | 5734 | 142 | 365 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 258 | 146751 | 197725 | 4726 | 96 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 442 | 15023 | 225773 | 47838 | 157 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 505 | 36063 | 35089 | 94315 | 2293 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 24327 | 76672 | 94323 | 20914 | 12092 | 905 | 164 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 3275 | 361946 | 48893 | 23714 | 1623 | 3317 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 4924 | 146668 | 156400 | 3624 | 115 | 1 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 13007 | 84333 | 75071 | 39219 | 23 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 1996 | 159860 | 134988 | 9142 | 2795 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 7399 | 27412 | 244105 | 10535 | 427 | 10 | 0 | 0 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 24 |
| 1989 | 10673 | 43756 | 23611 | 67102 | 1048 | 23 | 35 | 0 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 4 |
| 1990 | 16290 | 69073 | 30530 | 1772 | 4932 | 28 | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 11794 | 143967 | 40697 | 1163 | 17 | 107 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 36231 | 82605 | 115933 | 4063 | 97 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 12346 | 191714 | 163172 | 17474 | 170 | 1 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 19197 | 75840 | 254112 | 20271 | 2069 | 30 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 2118 | 231490 | 84163 | 67644 | 2539 | 199 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 22563 | 35010 | 413599 | 28996 | 10344 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 15260 | 114893 | 69948 | 106789 | 1700 | 425 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 2936 | 77065 | 162251 | 15801 | 20732 | 88 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 20814 | 57336 | 83205 | 46764 | 1905 | 2561 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 8472 | 320463 | 55818 | 24661 | 5703 | 321 | 201 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 1531 | 71284 | 521655 | 6483 | 1115 | 244 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2002 | 1120 | 21358 | 80304 | 243495 | 978 | 64 | 0 | 111 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 2937 | 7101 | 11014 | 31369 | 43849 | 13 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 3758 | 24613 | 21221 | 3967 | 14548 | 19811 | 5 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 8779 | 16730 | 18722 | 6181 | 1258 | 4826 | 4496 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2006 | 3229 | 118636 | 19862 | 8636 | 2634 | 823 | 1596 | 2520 | 6 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 8 |
| 2007 | 2045 | 19393 | 142509 | 5585 | 826 | 97 | 38 | 103 | 71 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 71 |
| 2008 | 3768 | 14623 | 25111 | 24195 | 243 | 46 | 134 | 2 | 30 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 34 |
| 2009 | 10468 | 10521 | 10601 | 11050 | 16522 | 79 | 50 | 46 | 0 | 1 | 4 | 0 | 0 | 0 | 0 | 0 | 5 |
| 2010 | 2930 | 102881 | 11872 | 5201 | 1125 | 2415 | 25 | 0 | 14 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 15 |
| 2011 | 3002 | 5858 | 49830 | 1817 | 806 | 105 | 224 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 2 |
| 2012 | 1319 | 3128 | 1973 | 14017 | 503 | 11 | 7 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 1285 | 11014 | 1898 | 494 | 2695 | 26 | 11 | 12 | 24 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 25 |
| 2014 | 3537 | 7272 | 7187 | 980 | 161 | 4185 | 2 | 14 | 7 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 8 |
| 2015 | 3820 | 26920 | 5225 | 1545 | 94 | 31 | 989 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 2 |

Table 13.2.9. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Numbers-at-age data (thousands) for IBC. Ages 0-7 and 8+ are used in the assessment.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ | $8+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 644461 | 254237 | 7686 | 183288 | 2365 | 592 | 118 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1966 | 1659093 | 548835 | 6546 | 1007 | 15861 | 755 | 25 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1967 | 298999 | 392510 | 5539 | 155 | 24 | 2264 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1968 | 11066 | 462938 | 81153 | 2029 | 46 | 19 | 738 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1969 | 70826 | 15178 | 1703617 | 98650 | 632 | 380 | 126 | 252 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1970 | 872884 | 170914 | 88509 | 485275 | 3967 | 153 | 61 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 323088 | 570391 | 9972 | 3390 | 6381 | 299 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 236664 | 238566 | 57010 | 1023 | 146 | 439 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 41332 | 117470 | 12402 | 11 | 11 | 196 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 576654 | 275266 | 32267 | 46862 | 4600 | 82 | 112 | 224 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 44317 | 594854 | 84620 | 4761 | 5203 | 141 | 5 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 |
| 1976 | 164982 | 66973 | 183064 | 29188 | 46 | 2946 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 103142 | 147019 | 29352 | 67628 | 2355 | 238 | 240 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 280592 | 125698 | 11330 | 809 | 1480 | 64 | 6 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 839615 | 125834 | 17671 | 1507 | 84 | 379 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 374315 | 281436 | 21820 | 8258 | 1291 | 54 | 86 | 7 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 1981 | 644910 | 99247 | 30358 | 4613 | 440 | 6 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 275003 | 174147 | 14740 | 13540 | 1810 | 464 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 488707 | 63818 | 14331 | 5134 | 2242 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 92587 | 98257 | 10644 | 4702 | 368 | 535 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 122079 | 11672 | 17826 | 1739 | 547 | 223 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 32696 | 40023 | 1831 | 802 | 103 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 8253 | 82226 | 3797 | 295 | 138 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 9280 | 3309 | 12819 | 2462 | 620 | 202 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 8914 | 2541 | 1751 | 2789 | 460 | 37 | 86 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 2996 | 7218 | 1986 | 359 | 1491 | 227 | 25 | 78 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 1991 | 116909 | 22493 | 3248 | 601 | 43 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 241702 | 80402 | 10971 | 356 | 27 | 3 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 124495 | 107664 | 13220 | 3214 | 82 | 9 | 0 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 69907 | 14349 | 9534 | 1011 | 160 | 7 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 198033 | 102560 | 2201 | 888 | 65 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 144469 | 10608 | 7453 | 3338 | 107 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 21694 | 45264 | 10935 | 4451 | 184 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 18983 | 9155 | 16337 | 2649 | 1490 | 63 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 69820 | 10780 | 4531 | 2932 | 344 | 166 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 4158 | 71419 | 21740 | 2085 | 186 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 1987 | 22946 | 35776 | 10127 | 35 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 49807 | 13889 | 4489 | 3638 | 504 | 27 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 4145 | 5983 | 2101 | 1285 | 1524 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 590 | 265 | 84 | 258 | 753 | 8 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 176 | 97 | 26 | 9 | 5 | 201 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 0 | 1772 | 716 | 241 | 47 | 46 | 74 | 108 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2007 | 1 | 27 | 218 | 6 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 12 | 82 | 280 | 180 | 52 | 18 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 15 | 36 | 97 | 48 | 19 | 6 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 4169 | 355 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 19 | 14 | 11 | 7 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 1 | 3 | 5 | 82 | 3 | 2 | 1 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 |
| 2014 | 0 | 0 | 20 | 6 | 12 | 67 | 2 | 2 | 1 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2015 | 0 | 6 | 9 | 1 | 3 | 12 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 13.2.10. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Mean weight at age data (kg) for total catch. Ages 0-7 and 8+ are used in the assessment.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 0.010 | 0.070 | 0.227 | 0.370 | 0.655 | 0.846 | 1.170 | 1.190 | 1.479 | 1.714 | 2.175 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1966 | 0.010 | 0.088 | 0.247 | 0.394 | 0.536 | 0.962 | 1.254 | 1.512 | 1.827 | 1.723 | 2.955 | 2.035 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1967 | 0.014 | 0.116 | 0.278 | 0.478 | 0.591 | 0.641 | 1.072 | 1.511 | 1.898 | 2.084 | 2.342 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1968 | 0.010 | 0.129 | 0.254 | 0.516 | 0.743 | 0.827 | 0.829 | 1.483 | 2.071 | 2.622 | 2.065 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1969 | 0.012 | 0.064 | 0.217 | 0.410 | 0.817 | 0.905 | 1.029 | 1.074 | 1.808 | 2.772 | 3.259 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1970 | 0.013 | 0.075 | 0.222 | 0.353 | 0.738 | 0.925 | 1.195 | 1.246 | 1.427 | 2.438 | 3.489 | 3.864 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1971 | 0.012 | 0.109 | 0.246 | 0.359 | 0.509 | 0.888 | 1.269 | 1.525 | 1.338 | 1.284 | 1.961 | 4.270 | 3.513 | 0.000 | 0.000 | 0.000 |
| 1972 | 0.025 | 0.117 | 0.242 | 0.383 | 0.503 | 0.585 | 0.987 | 1.380 | 1.967 | 1.979 | 1.618 | 2.861 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1973 | 0.043 | 0.118 | 0.239 | 0.369 | 0.578 | 0.611 | 0.648 | 1.044 | 1.378 | 2.658 | 1.603 | 1.988 | 2.123 | 0.000 | 0.000 | 0.000 |
| 1974 | 0.025 | 0.129 | 0.226 | 0.339 | 0.536 | 0.867 | 0.828 | 0.863 | 1.377 | 1.704 | 1.854 | 4.057 | 1.927 | 0.890 | 0.000 | 0.000 |
| 1975 | 0.023 | 0.105 | 0.240 | 0.353 | 0.442 | 0.678 | 1.190 | 1.077 | 1.031 | 1.564 | 2.188 | 2.764 | 0.000 | 3.318 | 0.000 | 0.000 |
| 1976 | 0.014 | 0.129 | 0.225 | 0.394 | 0.505 | 0.578 | 0.916 | 1.829 | 1.656 | 1.247 | 2.296 | 2.425 | 1.679 | 0.000 | 0.000 | 0.000 |
| 1977 | 0.020 | 0.111 | 0.238 | 0.339 | 0.586 | 0.612 | 0.787 | 1.160 | 1.715 | 1.971 | 1.490 | 2.067 | 0.000 | 3.898 | 0.000 | 0.000 |
| 1978 | 0.011 | 0.104 | 0.254 | 0.396 | 0.424 | 0.707 | 0.784 | 0.921 | 1.350 | 1.995 | 1.990 | 1.329 | 2.182 | 4.475 | 0.000 | 0.000 |
| 1979 | 0.009 | 0.093 | 0.287 | 0.417 | 0.611 | 0.669 | 0.931 | 1.241 | 1.320 | 1.453 | 2.505 | 1.575 | 1.233 | 1.580 | 0.000 | 0.000 |
| 1980 | 0.012 | 0.081 | 0.276 | 0.464 | 0.693 | 0.985 | 0.908 | 1.264 | 1.511 | 1.501 | 1.676 | 3.104 | 1.050 | 2.134 | 2.921 | 0.000 |
| 1981 | 0.009 | 0.060 | 0.264 | 0.445 | 0.726 | 1.055 | 1.222 | 1.195 | 1.545 | 1.672 | 1.531 | 1.515 | 2.982 | 4.273 | 1.896 | 0.000 |
| 1982 | 0.010 | 0.074 | 0.286 | 0.423 | 0.759 | 1.109 | 1.415 | 1.578 | 1.466 | 2.136 | 2.122 | 1.877 | 1.886 | 3.179 | 0.000 | 0.000 |
| 1983 | 0.011 | 0.132 | 0.303 | 0.431 | 0.612 | 0.904 | 1.211 | 1.191 | 1.630 | 1.460 | 1.449 | 1.972 | 2.853 | 4.689 | 0.000 | 0.000 |
| 1984 | 0.010 | 0.142 | 0.303 | 0.461 | 0.645 | 0.736 | 1.077 | 1.205 | 1.821 | 2.030 | 1.732 | 1.950 | 2.422 | 2.822 | 4.995 | 0.000 |
| 1985 | 0.010 | 0.148 | 0.296 | 0.466 | 0.649 | 0.835 | 0.934 | 1.344 | 1.638 | 2.097 | 2.109 | 2.061 | 2.555 | 2.471 | 2.721 | 4.139 |
| 1986 | 0.023 | 0.123 | 0.261 | 0.406 | 0.600 | 0.848 | 1.195 | 1.098 | 1.524 | 1.356 | 2.178 | 2.366 | 2.498 | 2.993 | 2.778 | 2.894 |
| 1987 | 0.010 | 0.125 | 0.264 | 0.405 | 0.594 | 0.974 | 1.215 | 1.322 | 1.260 | 1.358 | 1.870 | 2.132 | 2.609 | 2.450 | 2.768 | 2.638 |
| 1988 | 0.042 | 0.163 | 0.232 | 0.411 | 0.581 | 0.731 | 1.203 | 1.363 | 1.281 | 0.974 | 1.633 | 2.163 | 2.547 | 3.139 | 3.435 | 2.863 |
| 1989 | 0.036 | 0.200 | 0.282 | 0.367 | 0.590 | 0.770 | 0.935 | 1.259 | 1.586 | 1.507 | 1.034 | 1.534 | 2.431 | 2.559 | 2.307 | 0.980 |
| 1990 | 0.040 | 0.187 | 0.313 | 0.422 | 0.506 | 0.795 | 0.995 | 1.179 | 1.495 | 1.898 | 2.519 | 2.259 | 2.188 | 0.562 | 1.852 | 4.731 |
| 1991 | 0.030 | 0.175 | 0.308 | 0.454 | 0.574 | 0.644 | 0.959 | 1.136 | 1.313 | 1.701 | 2.163 | 2.012 | 1.622 | 1.070 | 1.208 | 2.888 |
| 1992 | 0.019 | 0.102 | 0.306 | 0.466 | 0.717 | 0.923 | 0.903 | 1.382 | 1.514 | 1.813 | 2.014 | 2.064 | 2.441 | 1.781 | 0.000 | 0.000 |
| 1993 | $0.010$ | 0.110 | 0.282 | 0.454 | 0.660 | 0.877 | 1.053 | 1.062 | 1.545 | 1.460 | 1.830 | 1.894 | 2.155 | 2.460 | 0.000 | 0.000 |
| 1994 | 0.018 | 0.121 | 0.247 | 0.435 | 0.599 | 0.846 | 1.240 | 1.274 | 1.289 | 1.573 | 2.060 | 2.070 | 2.834 | 2.403 | 2.523 | 0.000 |
| 1995 | 0.012 | 0.107 | 0.290 | 0.369 | 0.581 | 0.774 | 1.058 | 1.418 | 1.261 | 1.320 | 1.889 | 2.491 | 1.713 | 1.699 | 2.243 | 0.000 |
| 1996 | 0.022 | 0.126 | 0.241 | 0.382 | 0.484 | 0.746 | 0.847 | 0.825 | 1.616 | 1.538 | 1.433 | 1.830 | 2.358 | 2.636 | 3.433 | 0.000 |
| 1997 | 0.029 | 0.138 | 0.280 | 0.360 | 0.585 | 0.634 | 0.923 | 0.997 | 1.293 | 2.196 | 1.961 | 2.058 | 2.757 | 2.270 | 2.867 | 2.782 |
| 1998 | 0.027 | 0.153 | 0.255 | 0.396 | 0.444 | 0.665 | 0.777 | 1.041 | 1.109 | 1.251 | 2.373 | 2.334 | 1.656 | 2.433 | 2.085 | 2.509 |
| 1999 | 0.025 | 0.166 | 0.250 | 0.356 | 0.477 | 0.510 | 0.735 | 0.798 | 0.826 | 1.305 | 1.533 | 2.478 | 2.086 | 2.698 | 2.904 | 2.220 |
| 2000 | 0.052 | 0.121 | 0.256 | 0.355 | 0.480 | 0.605 | 0.656 | 1.033 | 0.973 | 1.529 | 1.911 | 2.323 | 2.365 | 2.310 | 3.595 | 1.843 |
| 2001 | 0.029 | 0.111 | 0.219 | 0.321 | 0.466 | 0.658 | 0.735 | 0.945 | 1.690 | 1.148 | 1.725 | 2.923 | 1.286 | 2.534 | 1.239 | 3.425 |
| 2002 | 0.017 | 0.109 | 0.255 | 0.311 | 0.527 | 0.703 | 0.829 | 0.818 | 1.279 | 1.945 | 1.798 | 1.839 | 2.352 | 2.762 | 0.000 | 0.000 |
| 2003 | 0.024 | 0.082 | 0.221 | 0.327 | 0.400 | 0.681 | 0.758 | 1.110 | 1.281 | 1.612 | 2.022 | 2.219 | 2.506 | 2.606 | 1.981 | 3.092 |
| 2004 | 0.039 | 0.139 | 0.238 | 0.378 | 0.395 | 0.440 | 0.686 | 0.926 | 1.184 | 1.602 | 1.753 | 2.605 | 2.170 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.054 | 0.160 | 0.271 | 0.364 | 0.495 | 0.479 | 0.522 | 0.925 | 1.054 | 1.373 | 1.847 | 2.750 | 2.545 | 2.309 | 3.431 | 0.000 |
| 2006 | 0.042 | 0.126 | 0.283 | 0.352 | 0.442 | 0.507 | 0.538 | 0.550 | 1.048 | 1.395 | 2.031 | 2.525 | 1.834 | 3.532 | 5.274 | 2.580 |
| 2007 | 0.042 | 0.159 | 0.227 | 0.407 | 0.478 | 0.538 | 0.657 | 0.700 | 0.745 | 0.902 | 2.272 | 0.971 | 1.712 | 2.348 | 4.244 | 0.000 |
| 2008 | 0.030 | 0.170 | 0.256 | 0.366 | 0.593 | 0.662 | 0.714 | 0.928 | 0.924 | 0.878 | 1.689 | 1.970 | 0.988 | 0.224 | 3.792 | 3.024 |
| 2009 | 0.048 | 0.175 | 0.305 | 0.323 | 0.388 | 0.677 | 0.799 | 0.839 | 1.308 | 1.318 | 1.025 | 1.045 | 1.150 | 3.091 | 2.115 | 0.000 |
| 2010 | 0.016 | 0.078 | 0.288 | 0.411 | 0.454 | 0.466 | 0.710 | 0.899 | 1.269 | 1.431 | 1.366 | 1.420 | 2.766 | 2.214 | 2.677 | 2.588 |
| 2011 | 0.017 | 0.140 | 0.260 | 0.399 | 0.434 | 0.466 | 0.534 | 0.661 | 0.864 | 0.558 | 1.484 | 1.787 | 1.593 | 0.000 | 0.000 | 0.000 |
| 2012 | 0.035 | 0.160 | 0.439 | 0.408 | 0.576 | 0.706 | 0.711 | 0.654 | 1.278 | 0.895 | 1.564 | 2.223 | 2.121 | 2.134 | 2.368 | 0.000 |
| 2013 | 0.034 | 0.172 | 0.425 | 0.599 | 0.487 | 0.727 | 0.854 | 0.796 | 0.758 | 1.085 | 1.842 | 2.191 | 2.607 | 1.810 | 2.512 | 0.000 |
| 2014 | 0.042 | 0.139 | 0.433 | 0.589 | 0.656 | 0.537 | 0.780 | 0.831 | 0.923 | 0.794 | 1.605 | 2.788 | 1.323 | 2.682 | 0.000 | 1.603 |
| 2015 | 0.031 | 0.145 | 0.417 | 0.561 | 0.752 | 0.698 | 0.631 | 0.685 | 0.970 | 0.725 | 0.715 | 0.719 | 1.448 | 2.954 | 0.000 | 0.000 |

Table 13.2.11. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Mean weight at age data (kg) for landings. Ages 0-7 and 8+ are used in the assessment.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 0.000 | 0.308 | 0.348 | 0.413 | 0.680 | 0.904 | 1.211 | 1.197 | 1.479 | 1.714 | 2.175 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1966 | 0.000 | 0.300 | 0.382 | 0.445 | 0.554 | 1.001 | 1.275 | 1.515 | 1.827 | 1.723 | 2.955 | 2.035 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1967 | 0.000 | 0.260 | 0.399 | 0.530 | 0.610 | 0.646 | 1.077 | 1.511 | 1.898 | 2.084 | 2.342 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1968 | 0.000 | 0.256 | 0.360 | 0.595 | 0.769 | 0.832 | 0.835 | 1.484 | 2.071 | 2.622 | 2.065 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1969 | 0.000 | 0.178 | 0.302 | 0.508 | 0.878 | 0.989 | 1.058 | 1.081 | 1.808 | 2.772 | 3.259 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1970 | 0.000 | 0.249 | 0.309 | 0.402 | 0.787 | 0.997 | 1.235 | 1.250 | 1.427 | 2.438 | 3.489 | 3.864 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1971 | 0.000 | 0.256 | 0.332 | 0.393 | 0.525 | 0.905 | 1.280 | 1.525 | 1.338 | 1.284 | 1.961 | 4.270 | 3.513 | 0.000 | 0.000 | 0.000 |
| 1972 | 0.000 | 0.243 | 0.325 | 0.415 | 0.518 | 0.587 | 0.989 | 1.380 | 1.967 | 1.979 | 1.618 | 2.861 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1973 | 0.000 | 0.228 | 0.310 | 0.400 | 0.596 | 0.621 | 0.649 | 1.044 | 1.378 | 2.658 | 1.603 | 1.988 | 2.123 | 0.000 | 0.000 | 0.000 |
| 1974 | 0.000 | 0.268 | 0.314 | 0.381 | 0.567 | 0.882 | 0.866 | 0.867 | 1.377 | 1.704 | 1.854 | 4.057 | 1.927 | 0.890 | 0.000 | 0.000 |
| 1975 | 0.000 | 0.254 | 0.336 | 0.400 | 0.476 | 0.683 | 1.193 | 1.077 | 1.031 | 1.564 | 2.188 | 2.764 | 0.000 | 3.318 | 0.000 | 0.000 |
| 1976 | 0.000 | 0.243 | 0.331 | 0.452 | 0.509 | 0.601 | 0.917 | 1.829 | 1.656 | 1.247 | 2.296 | 2.425 | 1.679 | 0.000 | 0.000 | 0.000 |
| 1977 | 0.000 | 0.272 | 0.344 | 0.381 | 0.595 | 0.625 | 0.800 | 1.160 | 1.715 | 1.971 | 1.490 | 2.067 | 0.000 | 3.898 | 0.000 | 0.000 |
| 1978 | 0.000 | 0.257 | 0.333 | 0.427 | 0.456 | 0.717 | 0.812 | 0.922 | 1.350 | 1.995 | 1.990 | 1.329 | 2.182 | 4.475 | 0.000 | 0.000 |
| 1979 | 0.000 | 0.262 | 0.348 | 0.447 | 0.620 | 0.675 | 0.932 | 1.241 | 1.320 | 1.453 | 2.505 | 1.575 | 1.233 | 1.580 | 0.000 | 0.000 |
| 1980 | 0.000 | 0.274 | 0.347 | 0.501 | 0.706 | 0.992 | 0.907 | 1.261 | 1.511 | 1.499 | 1.676 | 3.104 | 1.050 | 2.134 | 2.921 | 0.000 |
| 1981 | 0.000 | 0.334 | 0.364 | 0.503 | 0.734 | 1.056 | 1.222 | 1.195 | 1.545 | 1.672 | 1.531 | 1.515 | 2.982 | 4.273 | 1.896 | 0.000 |
| 1982 | 0.000 | 0.299 | 0.349 | 0.478 | 0.788 | 1.153 | 1.415 | 1.578 | 1.466 | 2.136 | 2.122 | 1.877 | 1.886 | 3.179 | 0.000 | 0.000 |
| 1983 | 0.000 | 0.320 | 0.375 | 0.464 | 0.624 | 0.914 | 1.242 | 1.191 | 1.630 | 1.460 | 1.449 | 1.972 | 2.853 | 4.689 | 0.000 | 0.000 |
| 1984 | 0.000 | 0.280 | 0.350 | 0.493 | 0.666 | 0.764 | 1.077 | 1.205 | 1.821 | 2.030 | 1.732 | 1.951 | 2.422 | 2.822 | 4.995 | 0.000 |
| 1985 | 0.000 | 0.279 | 0.348 | 0.478 | 0.651 | 0.844 | 0.935 | 1.344 | 1.638 | 2.097 | 2.109 | 2.061 | 2.555 | 2.471 | 2.721 | 4.139 |
| 1986 | 0.000 | 0.277 | 0.348 | 0.428 | 0.600 | 0.848 | 1.195 | 1.098 | 1.524 | 1.356 | 2.178 | 2.366 | 2.498 | 2.993 | 2.778 | 2.894 |
| 1987 | 0.000 | 0.265 | 0.335 | 0.440 | 0.603 | 0.974 | 1.215 | 1.322 | 1.260 | 1.358 | 1.870 | 2.132 | 2.609 | 2.450 | 2.768 | 2.638 |
| 1988 | 0.000 | 0.236 | 0.322 | 0.437 | 0.594 | 0.732 | 1.203 | 1.363 | 1.370 | 0.974 | 1.633 | 2.163 | 2.547 | 3.139 | 3.435 | 2.863 |
| 1989 | 0.000 | 0.319 | 0.356 | 0.413 | 0.602 | 0.769 | 0.934 | 1.256 | 1.579 | 1.507 | 1.025 | 1.534 | 2.431 | 2.559 | 2.307 | 0.980 |
| 1990 | 0.000 | 0.260 | 0.372 | 0.439 | 0.525 | 0.796 | 1.015 | 1.196 | 1.504 | 1.898 | 2.519 | 2.259 | 2.188 | 0.562 | 1.852 | 4.731 |
| 1991 | 0.000 | 0.269 | 0.363 | 0.462 | 0.576 | 0.645 | 0.959 | 1.136 | 1.313 | 1.701 | 2.163 | 2.012 | 1.622 | 1.070 | 1.208 | 2.888 |
| 1992 | 0.000 | 0.287 | 0.367 | 0.486 | 0.723 | 0.924 | 0.904 | 1.382 | 1.515 | 1.813 | 2.014 | 2.064 | 2.441 | 1.781 | 0.000 | 0.000 |
| 1993 | 0.000 | 0.293 | 0.372 | 0.484 | 0.666 | 0.878 | 1.053 | 1.067 | 1.545 | 1.460 | 1.830 | 1.894 | 2.155 | 2.460 | 0.000 | 0.000 |
| 1994 | 0.000 | 0.269 | 0.378 | 0.473 | 0.617 | 0.851 | 1.241 | 1.274 | 1.289 | 1.573 | 2.060 | 2.070 | 2.834 | 2.403 | 2.523 | 0.000 |
| 1995 | 0.000 | 0.316 | 0.400 | 0.424 | 0.600 | 0.782 | 1.058 | 1.418 | 1.261 | 1.320 | 1.889 | 2.491 | 1.713 | 1.699 | 2.243 | 0.000 |
| 1996 | 0.000 | 0.326 | 0.364 | 0.471 | 0.519 | 0.747 | 0.847 | 0.825 | 1.616 | 1.538 | 1.433 | 1.830 | 2.358 | 2.636 | 3.433 | 0.000 |
| 1997 | 0.000 | 0.344 | 0.410 | 0.418 | 0.615 | 0.641 | 0.923 | 0.997 | 1.293 | 2.196 | 1.961 | 2.058 | 2.757 | 2.270 | 2.867 | 2.782 |
| 1998 | 0.000 | 0.271 | 0.370 | 0.441 | 0.470 | 0.670 | 0.778 | 1.041 | 1.109 | 1.251 | 2.373 | 2.334 | 1.656 | 2.433 | 2.085 | 2.509 |
| 1999 | 0.000 | 0.297 | 0.349 | 0.422 | 0.490 | 0.523 | 0.746 | 0.798 | 0.826 | 1.305 | 1.533 | 2.478 | 2.086 | 2.698 | 2.904 | 2.220 |
| 2000 | 0.000 | 0.334 | 0.368 | 0.421 | 0.515 | 0.617 | 0.663 | 1.033 | 0.973 | 1.529 | 1.911 | 2.323 | 2.365 | 2.310 | 3.595 | 1.843 |
| 2001 | 0.000 | 0.379 | 0.352 | 0.448 | 0.483 | 0.675 | 0.735 | 0.946 | 1.695 | 1.148 | 1.725 | 2.923 | 1.286 | 2.534 | 1.239 | 3.425 |
| 2002 | 0.000 | 0.427 | 0.446 | 0.397 | 0.569 | 0.713 | 0.829 | 0.901 | 1.279 | 1.945 | 1.798 | 1.839 | 2.352 | 2.762 | 0.000 | 0.000 |
| 2003 | 0.000 | 0.283 | 0.377 | 0.464 | 0.441 | 0.684 | 0.759 | 1.110 | 1.281 | 1.612 | 2.022 | 2.219 | 2.506 | 2.606 | 1.981 | 3.092 |
| 2004 | 0.000 | 0.366 | 0.383 | 0.474 | 0.454 | 0.468 | 0.688 | 0.932 | 1.184 | 1.602 | 1.753 | 2.605 | 2.170 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.000 | 0.399 | 0.399 | 0.428 | 0.548 | 0.516 | 0.536 | 0.926 | 1.056 | 1.373 | 1.847 | 2.750 | 2.545 | 2.309 | 3.431 | 0.000 |
| 2006 | 0.000 | 0.392 | 0.386 | 0.418 | 0.493 | 0.546 | 0.574 | 0.583 | 1.093 | 1.431 | 2.109 | 2.643 | 1.926 | 3.592 | 5.292 | 2.709 |
| 2007 | 0.000 | 0.379 | 0.385 | 0.466 | 0.497 | 0.542 | 0.662 | 0.705 | 0.748 | 0.902 | 2.272 | 0.971 | 1.712 | 2.348 | 4.244 | 0.000 |
| 2008 | 0.000 | 0.357 | 0.408 | 0.414 | 0.607 | 0.668 | 0.754 | 0.931 | 0.935 | 0.879 | 1.703 | 1.970 | 0.988 | 0.224 | 3.792 | 3.024 |
| 2009 | 0.000 | 0.443 | 0.434 | 0.410 | 0.416 | 0.691 | 0.830 | 0.882 | 1.309 | 1.321 | 1.029 | 1.045 | 1.150 | 3.091 | 2.115 | 0.000 |
| 2010 | 0.000 | 0.278 | 0.473 | 0.457 | 0.471 | 0.476 | 0.721 | 0.899 | 1.364 | 1.431 | 1.366 | 1.420 | 2.766 | 2.214 | 2.677 | 2.588 |
| 2011 | 0.016 | 0.266 | 0.358 | 0.411 | 0.442 | 0.468 | 0.535 | 0.661 | 0.864 | 0.559 | 1.456 | 1.698 | 1.593 | 0.000 | 0.000 | 0.000 |
| 2012 | 0.000 | 0.358 | 0.525 | 0.445 | 0.606 | 0.707 | 0.712 | 0.654 | 1.279 | 0.895 | 1.564 | 2.223 | 2.121 | 2.134 | 2.368 | 0.000 |
| 2013 | 0.000 | 0.437 | 0.564 | 0.625 | 0.492 | 0.729 | 0.850 | 0.800 | 0.757 | 1.085 | 1.795 | 2.191 | 2.607 | 1.810 | 2.512 | 0.000 |
| 2014 | 0.000 | 0.311 | 0.510 | 0.654 | 0.662 | 0.557 | 0.781 | 0.834 | 0.932 | 0.794 | 1.605 | 2.788 | 1.323 | 2.682 | 0.000 | 1.603 |
| 2015 | 0.000 | 0.321 | 0.494 | 0.582 | 0.773 | 0.700 | 0.642 | 0.685 | 0.970 | 0.725 | 0.714 | 0.719 | 1.448 | 2.954 | 0.000 | 0.000 |

Table 13.2.12. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Mean weight at age data (kg) for discards. Ages 0-7 and 8+ are used in the assessment.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 0.062 | 0.131 | 0.203 | 0.335 | 0.607 | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1966 | 0.053 | 0.141 | 0.208 | 0.245 | 0.309 | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1967 | 0.043 | 0.170 | 0.210 | 0.273 | 0.306 | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1968 | 0.054 | 0.181 | 0.212 | 0.257 | 0.317 | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1969 | 0.049 | 0.129 | 0.216 | 0.238 | 0.300 | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1970 | 0.057 | 0.131 | 0.210 | 0.239 | 0.263 | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1971 | 0.052 | 0.135 | 0.202 | 0.244 | 0.264 | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1972 | 0.045 | 0.140 | 0.207 | 0.239 | 0.261 | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1973 | 0.051 | 0.135 | 0.201 | 0.237 | 0.263 | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1974 | 0.046 | 0.146 | 0.201 | 0.234 | 0.259 | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1975 | 0.041 | 0.126 | 0.201 | 0.257 | 0.275 | 0.348 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1976 | 0.053 | 0.172 | 0.198 | 0.239 | 0.291 | 0.337 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1977 | 0.062 | 0.191 | 0.198 | 0.220 | 0.306 | 0.347 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1978 | 0.042 | 0.175 | 0.199 | 0.222 | 0.225 | 0.265 | 0.284 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1979 | 0.037 | 0.128 | 0.221 | 0.245 | 0.259 | 0.314 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1980 | 0.051 | 0.147 | 0.232 | 0.276 | 0.325 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1981 | 0.074 | 0.160 | 0.199 | 0.296 | 0.621 | 0.727 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1982 | 0.055 | 0.194 | 0.247 | 0.265 | 0.289 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1983 | 0.066 | 0.184 | 0.237 | 0.343 | 0.458 | 0.711 | 0.792 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1984 | 0.047 | 0.160 | 0.245 | 0.315 | 0.309 | 0.290 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1985 | 0.040 | 0.154 | 0.221 | 0.271 | 0.356 | 0.423 | 0.353 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1986 | 0.057 | 0.140 | 0.185 | 0.246 | 0.337 | 0.329 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1987 | 0.026 | 0.160 | 0.201 | 0.227 | 0.286 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1988 | 0.072 | 0.167 | 0.172 | 0.239 | 0.256 | 0.352 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1989 | 0.054 | 0.188 | 0.229 | 0.266 | 0.336 | 0.708 | 0.844 | 0.000 | 2.572 | 0.000 | 3.048 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.047 | 0.189 | 0.229 | 0.248 | 0.264 | 0.290 | 0.333 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1991 | 0.059 | 0.179 | 0.238 | 0.341 | 0.464 | 0.480 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1992 | 0.043 | 0.136 | 0.246 | 0.282 | 0.345 | 0.000 | 0.592 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1993 | 0.028 | 0.139 | 0.237 | 0.287 | 0.355 | 0.369 | 0.000 | 0.430 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1994 | 0.042 | 0.130 | 0.212 | 0.273 | 0.310 | 0.304 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1995 | 0.044 | 0.132 | 0.250 | 0.276 | 0.356 | 0.384 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 0.047 | 0.133 | 0.218 | 0.279 | 0.297 | 0.335 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.060 | 0.159 | 0.250 | 0.286 | 0.322 | 0.374 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1998 | 0.075 | 0.159 | 0.232 | 0.293 | 0.317 | 0.391 | 0.428 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 0.047 | 0.182 | 0.217 | 0.273 | 0.308 | 0.304 | 0.227 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2000 | 0.049 | 0.129 | 0.245 | 0.278 | 0.316 | 0.355 | 0.292 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2001 | 0.049 | 0.115 | 0.206 | 0.300 | 0.301 | 0.300 | 0.000 | 0.411 | 0.416 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2002 | 0.044 | 0.125 | 0.223 | 0.267 | 0.334 | 0.382 | 0.000 | 0.358 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2003 | 0.042 | 0.124 | 0.223 | 0.261 | 0.327 | 0.536 | 0.630 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2004 | 0.039 | 0.135 | 0.218 | 0.263 | 0.299 | 0.330 | 0.639 | 0.650 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.054 | 0.150 | 0.232 | 0.273 | 0.318 | 0.301 | 0.342 | 0.499 | 0.493 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2006 | 0.042 | 0.121 | 0.231 | 0.265 | 0.279 | 0.274 | 0.217 | 0.164 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 0.042 | 0.146 | 0.195 | 0.291 | 0.314 | 0.358 | 0.375 | 0.356 | 0.368 | 0.400 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2008 | 0.030 | 0.166 | 0.217 | 0.262 | 0.365 | 0.456 | 0.317 | 0.454 | 0.427 | 0.596 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 0.048 | 0.162 | 0.250 | 0.248 | 0.282 | 0.394 | 0.315 | 0.357 | 0.366 | 0.409 | 0.452 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2010 | 0.016 | 0.076 | 0.209 | 0.303 | 0.307 | 0.315 | 0.350 | 0.523 | 0.284 | 0.000 | 0.000 | 1.445 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 0.017 | 0.135 | 0.227 | 0.297 | 0.310 | 0.352 | 0.351 | 0.000 | 0.000 | 0.000 | 2.027 | 2.215 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2012 | 0.035 | 0.143 | 0.295 | 0.271 | 0.286 | 0.406 | 0.353 | 0.392 | 0.633 | 0.488 | 0.316 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2013 | 0.034 | 0.148 | 0.243 | 0.362 | 0.345 | 0.498 | 1.355 | 0.533 | 0.842 | 0.000 | 2.113 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2014 | 0.042 | 0.133 | 0.298 | 0.336 | 0.394 | 0.340 | 0.572 | 0.617 | 0.475 | 0.885 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2015 | 0.031 | 0.141 | 0.261 | 0.347 | 0.377 | 0.411 | 0.407 | 0.634 | 0.634 | 0.000 | 1.082 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table 13.2.13. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Mean weight at age data (kg) for IBC. Ages 0-7 and 8+ are used in the assessment.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 0.010 | 0.040 | 0.180 | 0.302 | 0.400 | 0.420 | 0.440 | 0.500 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1966 | 0.010 | 0.040 | 0.180 | 0.302 | 0.400 | 0.420 | 0.440 | 0.500 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1967 | 0.010 | 0.040 | 0.180 | 0.302 | 0.400 | 0.420 | 0.440 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1968 | 0.010 | 0.040 | 0.180 | 0.302 | 0.400 | 0.420 | 0.440 | 0.500 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1969 | 0.010 | 0.040 | 0.180 | 0.302 | 0.400 | 0.420 | 0.440 | 0.500 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1970 | 0.010 | 0.040 | 0.180 | 0.302 | 0.400 | 0.420 | 0.440 | 0.500 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1971 | 0.010 | 0.040 | 0.180 | 0.302 | 0.400 | 0.420 | 0.440 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1972 | 0.023 | 0.067 | 0.136 | 0.255 | 0.288 | 0.231 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1973 | 0.035 | 0.068 | 0.141 | 0.246 | 0.327 | 0.396 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1974 | 0.022 | 0.058 | 0.150 | 0.260 | 0.359 | 0.579 | 0.277 | 0.447 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1975 | 0.020 | 0.039 | 0.173 | 0.275 | 0.267 | 0.413 | 0.585 | 0.000 | 0.585 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1976 | 0.012 | 0.046 | 0.181 | 0.304 | 0.473 | 0.360 | 0.725 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1977 | 0.013 | 0.042 | 0.184 | 0.307 | 0.490 | 0.352 | 0.442 | 1.234 | 1.315 | 1.319 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1978 | 0.011 | 0.040 | 0.174 | 0.286 | 0.372 | 0.473 | 0.411 | 0.456 | 1.315 | 0.000 | 1.400 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1979 | 0.009 | 0.039 | 0.177 | 0.285 | 0.384 | 0.461 | 0.735 | 1.234 | 1.315 | 0.000 | 1.400 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1980 | 0.012 | 0.039 | 0.176 | 0.268 | 0.623 | 0.722 | 1.102 | 1.591 | 0.000 | 1.796 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1981 | 0.009 | 0.040 | 0.176 | 0.371 | 0.467 | 0.858 | 1.200 | 1.234 | 1.315 | 1.319 | 1.400 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1982 | 0.010 | 0.040 | 0.206 | 0.379 | 0.636 | 0.751 | 1.225 | 1.233 | 1.315 | 1.319 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1983 | 0.008 | 0.047 | 0.173 | 0.428 | 0.584 | 1.006 | 1.225 | 1.234 | 1.315 | 1.319 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1984 | 0.009 | 0.045 | 0.211 | 0.414 | 0.626 | 0.751 | 1.225 | 1.234 | 1.315 | 1.319 | 1.400 | 1.400 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1985 | 0.009 | 0.043 | 0.186 | 0.371 | 0.550 | 0.563 | 0.565 | 1.234 | 1.315 | 1.319 | 1.400 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1986 | 0.010 | 0.040 | 0.186 | 0.375 | 0.626 | 1.259 | 1.225 | 1.234 | 1.315 | 1.319 | 1.400 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1987 | 0.006 | 0.038 | 0.258 | 0.442 | 0.908 | 1.171 | 1.225 | 1.234 | 1.315 | 1.319 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1988 | 0.018 | 0.077 | 0.196 | 0.274 | 0.455 | 0.549 | 1.225 | 1.234 | 1.315 | 1.319 | 1.400 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1989 | 0.015 | 0.165 | 0.251 | 0.347 | 0.670 | 0.923 | 1.065 | 1.492 | 1.315 | 0.000 | 1.400 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.005 | 0.104 | 0.229 | 0.506 | 0.609 | 0.842 | 0.829 | 0.796 | 0.956 | 1.319 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1991 | 0.027 | 0.058 | 0.206 | 0.357 | 0.472 | 0.477 | 1.225 | 1.234 | 1.315 | 1.319 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1992 | 0.015 | 0.059 | 0.217 | 0.422 | 0.552 | 0.615 | 0.548 | 1.234 | 0.621 | 0.820 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1993 | 0.008 | 0.053 | 0.206 | 0.399 | 0.521 | 0.578 | 1.225 | 0.582 | 1.315 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1994 | 0.011 | 0.055 | 0.155 | 0.435 | 0.595 | 0.698 | 0.490 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1995 | 0.012 | 0.045 | 0.193 | 0.285 | 0.387 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 0.018 | 0.077 | 0.136 | 0.162 | 0.264 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.007 | 0.076 | 0.149 | 0.309 | 0.419 | 0.601 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1998 | 0.020 | 0.075 | 0.166 | 0.291 | 0.351 | 0.453 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 0.018 | 0.064 | 0.177 | 0.304 | 0.416 | 0.309 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2000 | 0.058 | 0.070 | 0.113 | 0.176 | 0.370 | 0.203 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2001 | 0.014 | 0.086 | 0.133 | 0.110 | 0.353 | 0.431 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2002 | 0.016 | 0.064 | 0.178 | 0.283 | 0.374 | 0.431 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2003 | 0.012 | 0.031 | 0.056 | 0.231 | 0.326 | 0.339 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2004 | 0.000 | 0.116 | 0.183 | 0.255 | 0.276 | 0.446 | 0.539 | 0.840 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.000 | 0.107 | 0.187 | 0.239 | 0.268 | 0.287 | 0.598 | 0.619 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2006 | 0.000 | 0.127 | 0.232 | 0.273 | 0.273 | 0.280 | 0.283 | 0.286 | 0.287 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 0.035 | 0.141 | 0.192 | 0.290 | 0.315 | 0.370 | 0.427 | 0.342 | 0.368 | 0.400 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2008 | 0.042 | 0.146 | 0.291 | 0.388 | 0.454 | 0.526 | 0.414 | 0.406 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 0.047 | 0.180 | 0.252 | 0.247 | 0.279 | 0.410 | 0.417 | 0.413 | 0.400 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2010 | 0.000 | 0.080 | 0.244 | 0.310 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 0.016 | 0.316 | 0.324 | 0.350 | 0.367 | 0.443 | 0.460 | 0.493 | 0.589 | 0.385 | 0.000 | 1.331 | 1.624 | 0.000 | 0.000 | 0.000 |
| 2012 | 0.451 | 0.762 | 1.045 | 1.498 | 1.854 | 2.098 | 2.188 | 2.317 | 2.541 | 2.173 | 2.324 | 2.121 | 2.452 | 2.368 | 0.000 | 0.000 |
| 2013 | 0.000 | 0.437 | 0.564 | 0.626 | 0.492 | 0.729 | 0.850 | 0.800 | 0.757 | 1.085 | 1.795 | 2.191 | 2.607 | 1.810 | 2.512 | 0.000 |
| 2014 | 0.000 | 0.311 | 0.510 | 0.654 | 0.662 | 0.557 | 0.781 | 0.834 | 0.932 | 0.794 | 1.605 | 2.788 | 1.323 | 2.682 | 0.000 | 1.830 |
| 2015 | 0.000 | 0.321 | 0.494 | 0.582 | 0.773 | 0.700 | 0.642 | 0.685 | 0.970 | 0.725 | 0.714 | 0.719 | 1.448 | 2.954 | 0.000 | 0.000 |

Table 13.2.14. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Estimates of natural mortality from the most recent key run of SMS (ICES-WGSAM 2014).

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 1.466 | 1.508 | 0.843 | 0.529 | 0.466 | 0.321 | 0.268 | 0.243 | 0.219 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1966 | 1.466 | 1.508 | 0.843 | 0.529 | 0.466 | 0.321 | 0.268 | 0.243 | 0.219 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1967 | 1.466 | 1.508 | 0.843 | 0.529 | 0.466 | 0.321 | 0.268 | 0.243 | 0.219 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1968 | 1.466 | 1.508 | 0.843 | 0.529 | 0.466 | 0.321 | 0.268 | 0.243 | 0.219 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1969 | 1.466 | 1.508 | 0.843 | 0.529 | 0.466 | 0.321 | 0.268 | 0.243 | 0.219 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1970 | 1.466 | 1.508 | 0.843 | 0.529 | 0.466 | 0.321 | 0.268 | 0.243 | 0.219 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1971 | 1.466 | 1.508 | 0.843 | 0.529 | 0.466 | 0.321 | 0.268 | 0.243 | 0.219 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1972 | 1.466 | 1.508 | 0.843 | 0.529 | 0.466 | 0.321 | 0.268 | 0.243 | 0.219 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1973 | 1.466 | 1.508 | 0.843 | 0.529 | 0.466 | 0.321 | 0.268 | 0.243 | 0.219 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1974 | 1.466 | 1.508 | 0.843 | 0.529 | 0.466 | 0.321 | 0.268 | 0.243 | 0.219 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1975 | 1.511 | 1.528 | 0.820 | 0.511 | 0.441 | 0.314 | 0.264 | 0.238 | 0.217 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1976 | 1.551 | 1.547 | 0.798 | 0.494 | 0.417 | 0.306 | 0.261 | 0.233 | 0.215 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1977 | 1.583 | 1.565 | 0.775 | 0.477 | 0.393 | 0.297 | 0.257 | 0.230 | 0.212 | 0.206 | 0.200 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 1978 | 1.605 | 1.578 | 0.753 | 0.462 | 0.372 | 0.287 | 0.252 | 0.226 | 0.210 | 0.206 | 0.200 | 0.232 | 0.232 | 0.232 | 0.232 | 0.232 |
| 1979 | 1.618 | 1.583 | 0.731 | 0.447 | 0.351 | 0.277 | 0.246 | 0.222 | 0.208 | 0.205 | 0.200 | 0.231 | 0.231 | 0.231 | 0.231 | 0.231 |
| 1980 | 1.624 | 1.579 | 0.708 | 0.433 | 0.333 | 0.269 | 0.240 | 0.219 | 0.207 | 0.205 | 0.200 | 0.230 | 0.230 | 0.230 | 0.230 | 0.230 |
| 1981 | 1.622 | 1.566 | 0.685 | 0.420 | 0.318 | 0.261 | 0.235 | 0.217 | 0.205 | 0.205 | 0.200 | 0.228 | 0.228 | 0.228 | 0.228 | 0.228 |
| 1982 | 1.616 | 1.539 | 0.662 | 0.409 | 0.306 | 0.256 | 0.230 | 0.215 | 0.204 | 0.204 | 0.200 | 0.226 | 0.226 | 0.226 | 0.226 | 0.226 |
| 1983 | 1.609 | 1.500 | 0.637 | 0.398 | 0.297 | 0.253 | 0.226 | 0.214 | 0.203 | 0.204 | 0.200 | 0.224 | 0.224 | 0.224 | 0.224 | 0.224 |
| 1984 | 1.603 | 1.452 | 0.612 | 0.387 | 0.291 | 0.251 | 0.224 | 0.213 | 0.202 | 0.204 | 0.200 | 0.222 | 0.222 | 0.222 | 0.222 | 0.222 |
| 1985 | 1.597 | 1.404 | 0.589 | 0.376 | 0.287 | 0.249 | 0.222 | 0.212 | 0.202 | 0.203 | 0.200 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 |
| 1986 | 1.589 | 1.358 | 0.567 | 0.366 | 0.284 | 0.248 | 0.221 | 0.211 | 0.201 | 0.203 | 0.200 | 0.217 | 0.217 | 0.217 | 0.217 | 0.217 |
| 1987 | 1.577 | 1.318 | 0.545 | 0.357 | 0.282 | 0.247 | 0.220 | 0.210 | 0.201 | 0.202 | 0.200 | 0.215 | 0.215 | 0.215 | 0.215 | 0.215 |
| 1988 | 1.555 | 1.285 | 0.525 | 0.349 | 0.281 | 0.246 | 0.220 | 0.209 | 0.201 | 0.202 | 0.200 | 0.213 | 0.213 | 0.213 | 0.213 | 0.213 |
| 1989 | 1.525 | 1.257 | 0.507 | 0.341 | 0.281 | 0.246 | 0.220 | 0.209 | 0.201 | 0.201 | 0.200 | 0.211 | 0.211 | 0.211 | 0.211 | 0.211 |
| 1990 | 1.487 | 1.234 | 0.489 | 0.333 | 0.280 | 0.245 | 0.220 | 0.209 | 0.201 | 0.201 | 0.200 | 0.210 | 0.210 | 0.210 | 0.210 | 0.210 |
| 1991 | 1.444 | 1.215 | 0.472 | 0.325 | 0.280 | 0.244 | 0.220 | 0.209 | 0.201 | 0.201 | 0.200 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 |
| 1992 | 1.401 | 1.203 | 0.458 | 0.319 | 0.279 | 0.243 | 0.220 | 0.209 | 0.202 | 0.200 | 0.200 | 0.207 | 0.207 | 0.207 | 0.207 | 0.207 |
| 1993 | 1.364 | 1.196 | 0.446 | 0.313 | 0.278 | 0.241 | 0.220 | 0.209 | 0.202 | 0.200 | 0.200 | 0.207 | 0.207 | 0.207 | 0.207 | 0.207 |
| 1994 | 1.333 | 1.194 | 0.437 | 0.309 | 0.278 | 0.240 | 0.220 | 0.210 | 0.203 | 0.200 | 0.200 | 0.206 | 0.206 | 0.206 | 0.206 | 0.206 |
| 1995 | 1.311 | 1.197 | 0.430 | 0.306 | 0.277 | 0.240 | 0.220 | 0.210 | 0.203 | 0.200 | 0.200 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 |
| 1996 | 1.298 | 1.202 | 0.425 | 0.305 | 0.277 | 0.240 | 0.221 | 0.211 | 0.204 | 0.200 | 0.200 | 0.204 | 0.204 | 0.204 | 0.204 | 0.204 |
| 1997 | 1.292 | 1.211 | 0.422 | 0.305 | 0.276 | 0.241 | 0.221 | 0.211 | 0.205 | 0.200 | 0.200 | 0.204 | 0.204 | 0.204 | 0.204 | 0.204 |
| 1998 | 1.292 | 1.222 | 0.421 | 0.305 | 0.275 | 0.241 | 0.222 | 0.211 | 0.205 | 0.200 | 0.200 | 0.203 | 0.203 | 0.203 | 0.203 | 0.203 |
| 1999 | 1.299 | 1.238 | 0.420 | 0.306 | 0.274 | 0.241 | 0.223 | 0.211 | 0.206 | 0.201 | 0.200 | 0.203 | 0.203 | 0.203 | 0.203 | 0.203 |
| 2000 | 1.310 | 1.260 | 0.421 | 0.306 | 0.272 | 0.240 | 0.223 | 0.211 | 0.206 | 0.201 | 0.200 | 0.203 | 0.203 | 0.203 | 0.203 | 0.203 |
| 2001 | 1.320 | 1.289 | 0.424 | 0.306 | 0.271 | 0.241 | 0.224 | 0.211 | 0.206 | 0.201 | 0.200 | 0.203 | 0.203 | 0.203 | 0.203 | 0.203 |
| 2002 | 1.322 | 1.320 | 0.430 | 0.306 | 0.271 | 0.241 | 0.225 | 0.212 | 0.205 | 0.201 | 0.200 | 0.204 | 0.204 | 0.204 | 0.204 | 0.204 |
| 2003 | 1.313 | 1.346 | 0.437 | 0.306 | 0.271 | 0.243 | 0.226 | 0.213 | 0.205 | 0.202 | 0.200 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 |
| 2004 | 1.291 | 1.362 | 0.446 | 0.305 | 0.271 | 0.246 | 0.228 | 0.214 | 0.205 | 0.202 | 0.200 | 0.206 | 0.206 | 0.206 | 0.206 | 0.206 |
| 2005 | 1.259 | 1.361 | 0.456 | 0.303 | 0.272 | 0.248 | 0.229 | 0.215 | 0.205 | 0.202 | 0.200 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 |
| 2006 | 1.222 | 1.346 | 0.466 | 0.303 | 0.274 | 0.252 | 0.231 | 0.216 | 0.205 | 0.203 | 0.201 | 0.209 | 0.209 | 0.209 | 0.209 | 0.209 |
| 2007 | 1.183 | 1.320 | 0.479 | 0.304 | 0.277 | 0.256 | 0.232 | 0.216 | 0.205 | 0.203 | 0.201 | 0.212 | 0.212 | 0.212 | 0.212 | 0.212 |
| 2008 | 1.147 | 1.288 | 0.492 | 0.308 | 0.283 | 0.263 | 0.233 | 0.216 | 0.204 | 0.203 | 0.201 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 |
| 2009 | 1.115 | 1.257 | 0.507 | 0.313 | 0.290 | 0.273 | 0.235 | 0.216 | 0.204 | 0.202 | 0.201 | 0.216 | 0.216 | 0.216 | 0.216 | 0.216 |
| 2010 | 1.089 | 1.231 | 0.523 | 0.321 | 0.300 | 0.286 | 0.238 | 0.216 | 0.203 | 0.202 | 0.201 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 |
| 2011 | 1.067 | 1.212 | 0.541 | 0.332 | 0.312 | 0.301 | 0.242 | 0.217 | 0.202 | 0.201 | 0.201 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 |
| 2012 | 1.046 | 1.199 | 0.561 | 0.344 | 0.326 | 0.319 | 0.247 | 0.218 | 0.201 | 0.201 | 0.201 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 |
| 2013 | 1.024 | 1.188 | 0.581 | 0.357 | 0.340 | 0.337 | 0.252 | 0.219 | 0.201 | 0.200 | 0.201 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 |
| 2014 | 1.024 | 1.188 | 0.581 | 0.357 | 0.340 | 0.337 | 0.252 | 0.219 | 0.201 | 0.200 | 0.201 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 |
| 2015 | 1.024 | 1.188 | 0.581 | 0.357 | 0.340 | 0.337 | 0.252 | 0.219 | 0.201 | 0.200 | 0.201 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 |

Table 13.2.15. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Data available for calibration of the assessment. Only those data used in the final assessment are shown here.

| North Sea IBTS Q1 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | 2016 |  |  |  |  |
| 1 | 1 | 0.00 | 0.25 |  |  |
| 1 | 5 |  |  |  |  |
| 100 | 302.278 | 403.079 | 89.463 | 116.447 | 13.182 |
| 100 | 1072.285 | 221.275 | 127.77 | 20.41 | 20.9 |
| 100 | 230.968 | 833.257 | 107.598 | 32.317 | 3.575 |
| 100 | 573.023 | 266.912 | 303.546 | 17.888 | 6.49 |
| 100 | 912.559 | 328.062 | 45.201 | 58.262 | 4.345 |
| 100 | 101.691 | 677.641 | 97.149 | 12.684 | 13.965 |
| 100 | 219.06 | 97.372 | 273.008 | 16.604 | 2.114 |
| 100 | 217.448 | 139.114 | 32.997 | 50.367 | 3.163 |
| 100 | 680.231 | 134.076 | 25.032 | 4.26 | 8.476 |
| 100 | 1141.396 | 331.044 | 17.035 | 3.026 | 0.664 |
| 100 | 1242.121 | 519.521 | 152.384 | 8.848 | 1.076 |
| 100 | 227.919 | 491.051 | 97.656 | 23.308 | 1.566 |
| 100 | 1355.485 | 201.069 | 176.165 | 24.354 | 5.286 |
| 100 | 267.411 | 813.268 | 65.869 | 46.691 | 7.734 |
| 100 | 848.966 | 354.766 | 466.823 | 24.987 | 15.238 |
| 100 | 357.597 | 420.926 | 103.531 | 112.632 | 8.758 |
| 100 | 211.139 | 222.907 | 127.063 | 48.217 | 36.649 |
| 100 | 3734.174 | 107.12 | 48.609 | 24.497 | 15.58 |
| 100 | 901.378 | 2216.722 | 75.408 | 14.506 | 7.244 |
| 100 | 57.312 | 473.628 | 1309.589 | 9.179 | 6.886 |
| 100 | 89.991 | 39.267 | 241.529 | 532.024 | 5.354 |
| 100 | 71.745 | 79.256 | 36.962 | 176.352 | 324.91 |
| 100 | 70.189 | 51.885 | 38.458 | 14.057 | 54.576 |
| 100 | 1158.194 | 46.081 | 28.477 | 9.896 | 4.837 |
| 100 | 109.44 | 963.393 | 35.962 | 14.956 | 3.019 |
| 100 | 61.357 | 107.39 | 241.221 | 14.886 | 1.592 |
| 100 | 75.068 | 141.444 | 102.986 | 135.595 | 2.528 |
| 100 | 674.962 | 71.132 | 68.015 | 51.48 | 90.942 |
| 100 | 46.068 | 781.507 | 101.666 | 35.942 | 47.87 |
| 100 | 14.006 | 66.409 | 390.588 | 21.18 | 15.108 |
| 100 | 58.227 | 24.55 | 32.549 | 93.814 | 6.488 |
| 100 | 24.066 | 104.024 | 18.339 | 49.978 | 126.068 |
| 100 | 388.205 | 32.597 | 29.955 | 3.879 | 9.103 |
| 100 | 104.434 | 282.372 | 15.419 | 11.375 | 1.905 |

Table 13.2.15. (cont.) Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Data available for calibration of the assessment. Only those data used in the final assessment are shown here. Note that these pertain to the assessment produced by IBPHaddock in November 2016, so the Q3 2016 data is included (as it was used in that assessment).

| North Sea IBTS Q3 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 2016 |  |  |  |  |  |
| 1 | 1 | 0.50 | 0.75 |  |  |  |
| 0 | 5 |  |  |  |  |  |
| 100 | 718.479 | 233.55 | 22.921 | 2.842 | 0.507 | 1.561 |
| 100 | 2741.14 | 595.235 | 189.015 | 10.529 | 1.583 | 0.396 |
| 100 | 577.382 | 605.99 | 140.146 | 37.604 | 2.36 | 0.372 |
| 100 | 1781.191 | 195.331 | 262.643 | 32.423 | 8.383 | 0.381 |
| 100 | 520.855 | 1019.607 | 106.642 | 97.383 | 8.06 | 3.131 |
| 100 | 627.502 | 247.469 | 428.471 | 30.426 | 20.215 | 2.649 |
| 100 | 195.255 | 347.567 | 123.793 | 149.048 | 6.672 | 5.282 |
| 100 | 276.401 | 257.14 | 164.853 | 53.69 | 42.66 | 3.093 |
| 100 | 6904.537 | 176.457 | 94.108 | 47.947 | 13.268 | 9.904 |
| 100 | 1092.83 | 2511.127 | 44.361 | 19.494 | 10.29 | 4.276 |
| 100 | 34.751 | 360.531 | 1100.248 | 30.305 | 6.377 | 3.653 |
| 100 | 138.204 | 49.504 | 223.792 | 583.061 | 10.079 | 2.601 |
| 100 | 163.924 | 69.356 | 31.171 | 199.252 | 368.656 | 2.942 |
| 100 | 183.977 | 69.539 | 40.556 | 23.119 | 82.685 | 154.82 |
| 100 | 1412.973 | 67.605 | 45.54 | 16.254 | 9.845 | 37.095 |
| 100 | 191.608 | 547.284 | 27.543 | 11.709 | 3.612 | 3.352 |
| 100 | 111.475 | 149.743 | 385.791 | 10.354 | 5.35 | 1.126 |
| 100 | 126.428 | 86.627 | 89.934 | 174.968 | 5.206 | 2.253 |
| 100 | 909.334 | 77.703 | 79.994 | 38.131 | 73.972 | 1.643 |
| 100 | 30.294 | 557.39 | 59.017 | 34.214 | 25.186 | 53.33 |
| 100 | 30.64 | 77.035 | 344.508 | 27.159 | 12.209 | 9.196 |
| 100 | 68.068 | 31.515 | 40.248 | 132.237 | 7.344 | 4.397 |
| 100 | 86.249 | 58.345 | 25.17 | 18.291 | 82.779 | 2.515 |
| 100 | 747.522 | 48.207 | 58.51 | 5.216 | 9.093 | 51.625 |
| 100 | 104.274 | 463.428 | 22.807 | 15.993 | 1.662 | 2.307 |
| 100 | 351.819 | 94.546 | 219.874 | 8.057 | 3.669 | 0.400 |

Table 13.3.1. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. TSA final assessment: Model settings. $\omega$ is a multiplier on the permitted variance of the estimated value: a higher setting for $\omega$ indicates greater downweighting of that value in the overall assessment.

| Landings | Ages | 0-8+ |
| :---: | :---: | :---: |
|  | Years | 1972-2015 |
| Discards | Ages | 0-8+ |
|  | Years | 1972, 1978-2015 |
| Industrial bycatch | Ages | 0-8+ |
|  | Years | 1972, 1978-2015 |
| Survey: NS IBTS Q1 | Ages | 1-5 |
|  | Years | 1983-2016 |
| Survey: NS IBTS Q3 | Ages | 0-5 |
|  | Years | 1991-2016 |
| Maturity |  | Knife-edge at age 3 (interim measure) |
| Natural mortality |  | Age- and time-varying from North Sea SMS key runs |
| Catch weights |  | Catch abundance-weighted average of North Sea and West of Scotland catch weights |
| Stock weights |  | Set equal to catch weights (interim measure) |
| Large year-classes ( $\lambda=5$ ) |  | 1974, 1979, 1999 |
| Age-dependent F variability |  | $H(a)=(2,1,1,1,1,1,1,1,1)$ |
| F plateau |  | $a_{m}=5$ |
| Measurement-error multiplier for landings |  | $B_{\text {landings }}(a)=(*, 3.7,1.3,1,1.1,1.4,1.6,2.7,2.8)$ |
| Measurement-error multiplier for discards+bycatch |  | $B_{\text {discards }}(a)=(2.0,1.7,1,1.5,1.8,2.4, *, *, *)$ |
| Downweighted landings outliers |  | 1996, age $7(\omega=3)$ |
| Downweighted discards+bycatch outliers |  | 1982, age 5; 2012, age 2 ( $\omega=3$ for both) |
| Downweighted survey outliers |  | NS IBST Q1: 2011, age 5 ( $\omega=3$ ) |

Table 13.3.2. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. TSA final assessment: Parameter estimates.

|  | estimat <br> e | lower bound | upper <br> bound | Estimate d | on bound |
| :---: | :---: | :---: | :---: | :---: | :---: |
| F age 0 | 0.0351 | 0.005 | 0.1 | TRUE | FALSE |
| $F$ age 1 | 0.0795 | 0.05 | 0.15 | TRUE | FALSE |
| $F$ age 2 | 0.8515 | 0.6 | 1 | TRUE | FALSE |
| $F$ age 7 | 1.306 | 1 | 1.4 | TRUE | FALSE |
| sd F | 0.1621 | 0.01 | 0.2 | TRUE | FALSE |
| sd U | 0.069 | 0.01 | 0.15 | TRUE | FALSE |
| sd V | 0.1446 | 0.01 | 0.2 | TRUE | FALSE |
| sd Y | 0.1544 | 0.01 | 0.25 | TRUE | FALSE |
| cv landings | 0.1492 | 0.1 | 0.3 | TRUE | FALSE |
| cv discards | 0.2874 | 0.2 | 0.4 | TRUE | FALSE |
| log mean recruitment at start | 7.2937 | 7 | 9 | TRUE | FALSE |
| sd of random walk | 0.0713 | 0 | 0.25 | TRUE | FALSE |
| recruitment cv | 0.5236 | 0.3 | 0.6 | TRUE | FALSE |
| discards sd transitory | 0.0091 | 0 | 0.35 | TRUE | FALSE |
| discards sd persistent | 0.3378 | 0.25 | 0.5 | TRUE | FALSE |
| NSQ1 selection age 1 | 0.2527 | 0.1 | 0.3 | TRUE | FALSE |
| NSQ1 selection age 2 | 0.6436 | 0.4 | 0.8 | TRUE | FALSE |
| NSQ1 selection age 3 | 0.7099 | 0.6 | 0.9 | TRUE | FALSE |
| NSQ1 selection age 4 | 0.5823 | 0.4 | 0.8 | TRUE | FALSE |
| NSQ1 selection age 5 | 0.4609 | 0.4 | 0.8 | TRUE | FALSE |
| NSQ1 sigma | 0.3469 | 0.1 | 0.4 | TRUE | FALSE |
| NSQ1 eta | 0.2138 | 0.1 | 0.8 | TRUE | FALSE |
| NSQ1 omega | 0.0898 | 0 | 0.3 | TRUE | FALSE |
| NSQ1 beta | 0 | 0 | 0.1 | FALSE | TRUE |
| NSQ3 selection age 0 | 0.2248 | 0.1 | 0.4 | TRUE | FALSE |
| NSQ3 selection age 1 | 0.3717 | 0.2 | 0.6 | TRUE | FALSE |
| NSQ3 selection age 2 | 0.5643 | 0.2 | 0.8 | TRUE | FALSE |
| NSQ3 selection age 3 | 0.4903 | 0.2 | 0.8 | TRUE | FALSE |
| NSQ3 selection age 4 | 0.3858 | 0.2 | 0.8 | TRUE | FALSE |
| NSQ3 selection age 5 | 0.3477 | 0.2 | 0.8 | TRUE | FALSE |
| NSQ3 sigma | 0.2382 | 0.1 | 0.4 | TRUE | FALSE |
| NSQ3 eta | 0.083 | 0 | 0.3 | TRUE | FALSE |
| NSQ3 omega | 0.0925 | 0 | 0.3 | TRUE | FALSE |
| NSQ3 beta | 0 | 0 | 0.1 | FALSE | TRUE |

Table 13.3.3. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Estimates of fishing mortality at age from the final TSA assessment. Estimates refer to the full year (January - December) except for age 0 , for which the mortality rate given refers to the second half-year only (July December). The 2016 estimates ( ${ }^{*}$ ) are TSA forecasts.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | Mean F(2-4) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 0.033 | 0.077 | 0.607 | 0.998 | 0.953 | 0.917 | 1.011 | 1.055 | 0.997 | 0.852 |
| 1973 | 0.029 | 0.079 | 0.606 | 0.872 | 0.847 | 0.893 | 0.995 | 1.037 | 1.101 | 0.775 |
| 1974 | 0.027 | 0.080 | 0.613 | 0.720 | 0.835 | 0.759 | 0.893 | 0.961 | 0.965 | 0.723 |
| 1975 | 0.030 | 0.081 | 0.696 | 0.883 | 0.965 | 0.928 | 1.089 | 1.080 | 1.069 | 0.848 |
| 1976 | 0.028 | 0.083 | 0.554 | 0.966 | 0.857 | 1.035 | 0.969 | 0.996 | 1.001 | 0.792 |
| 1977 | 0.027 | 0.091 | 0.602 | 0.748 | 1.057 | 0.965 | 0.969 | 0.944 | 0.968 | 0.802 |
| 1978 | 0.023 | 0.113 | 0.647 | 0.943 | 1.079 | 1.072 | 1.065 | 1.075 | 1.111 | 0.890 |
| 1979 | 0.027 | 0.095 | 0.688 | 1.049 | 0.990 | 1.012 | 1.027 | 1.042 | 1.048 | 0.909 |
| 1980 | 0.029 | 0.078 | 0.489 | 1.045 | 1.101 | 0.798 | 0.912 | 0.958 | 0.959 | 0.878 |
| 1981 | 0.025 | 0.069 | 0.321 | 0.779 | 0.893 | 0.740 | 0.475 | 0.729 | 0.698 | 0.664 |
| 1982 | 0.018 | 0.069 | 0.372 | 0.579 | 0.695 | 0.591 | 0.600 | 0.700 | 0.624 | 0.548 |
| 1983 | 0.017 | 0.079 | 0.442 | 0.837 | 0.852 | 0.893 | 0.750 | 0.746 | 0.761 | 0.710 |
| 1984 | 0.020 | 0.108 | 0.483 | 0.937 | 1.075 | 0.818 | 0.829 | 0.802 | 0.802 | 0.832 |
| 1985 | 0.020 | 0.108 | 0.441 | 0.909 | 1.010 | 0.864 | 0.821 | 0.768 | 0.772 | 0.787 |
| 1986 | 0.016 | 0.112 | 0.628 | 0.922 | 1.102 | 0.820 | 0.684 | 0.685 | 0.727 | 0.884 |
| 1987 | 0.021 | 0.091 | 0.720 | 1.002 | 0.948 | 0.877 | 0.882 | 0.814 | 0.790 | 0.890 |
| 1988 | 0.020 | 0.109 | 0.578 | 1.153 | 1.087 | 0.939 | 0.855 | 0.780 | 0.817 | 0.939 |
| 1989 | 0.018 | 0.111 | 0.625 | 0.939 | 1.102 | 0.870 | 0.845 | 0.776 | 0.781 | 0.888 |
| 1990 | 0.015 | 0.108 | 0.708 | 0.967 | 0.982 | 0.858 | 0.733 | 0.686 | 0.702 | 0.886 |
| 1991 | 0.016 | 0.150 | 0.686 | 1.014 | 0.925 | 0.785 | 0.780 | 0.734 | 0.701 | 0.875 |
| 1992 | 0.018 | 0.114 | 0.630 | 0.993 | 0.997 | 0.667 | 0.857 | 0.699 | 0.722 | 0.874 |
| 1993 | 0.020 | 0.149 | 0.784 | 0.997 | 1.014 | 0.969 | 0.841 | 0.817 | 0.833 | 0.932 |
| 1994 | 0.014 | 0.118 | 0.721 | 1.032 | 0.981 | 1.031 | 0.976 | 0.901 | 0.827 | 0.911 |
| 1995 | 0.018 | 0.095 | 0.588 | 0.922 | 0.948 | 0.827 | 0.921 | 0.712 | 0.708 | 0.820 |
| 1996 | 0.016 | 0.093 | 0.521 | 0.879 | 1.014 | 0.979 | 0.965 | 0.704 | 0.700 | 0.805 |
| 1997 | 0.013 | 0.108 | 0.485 | 0.645 | 0.758 | 0.902 | 0.792 | 0.607 | 0.594 | 0.629 |
| 1998 | 0.013 | 0.133 | 0.619 | 0.702 | 0.883 | 0.834 | 0.809 | 0.617 | 0.604 | 0.735 |
| 1999 | 0.011 | 0.119 | 0.675 | 0.926 | 0.871 | 1.092 | 0.892 | 0.677 | 0.649 | 0.824 |
| 2000 | 0.011 | 0.096 | 0.732 | 0.974 | 0.972 | 0.844 | 0.875 | 0.615 | 0.592 | 0.893 |
| 2001 | 0.010 | 0.078 | 0.419 | 0.700 | 0.723 | 0.681 | 0.612 | 0.435 | 0.422 | 0.614 |
| 2002 | 0.006 | 0.095 | 0.281 | 0.380 | 0.499 | 0.484 | 0.433 | 0.294 | 0.292 | 0.387 |
| 2003 | 0.004 | 0.045 | 0.218 | 0.235 | 0.279 | 0.344 | 0.288 | 0.188 | 0.184 | 0.244 |
| 2004 | 0.004 | 0.050 | 0.216 | 0.252 | 0.261 | 0.317 | 0.254 | 0.161 | 0.158 | 0.243 |
| 2005 | 0.003 | 0.055 | 0.280 | 0.356 | 0.284 | 0.337 | 0.311 | 0.172 | 0.168 | 0.307 |
| 2006 | 0.005 | 0.052 | 0.426 | 0.534 | 0.551 | 0.533 | 0.399 | 0.262 | 0.225 | 0.503 |
| 2007 | 0.005 | 0.054 | 0.237 | 0.517 | 0.514 | 0.499 | 0.386 | 0.220 | 0.215 | 0.423 |
| 2008 | 0.003 | 0.036 | 0.181 | 0.230 | 0.338 | 0.318 | 0.263 | 0.143 | 0.142 | 0.250 |
| 2009 | 0.002 | 0.030 | 0.128 | 0.194 | 0.263 | 0.251 | 0.188 | 0.110 | 0.104 | 0.195 |
| 2010 | 0.003 | 0.031 | 0.160 | 0.234 | 0.229 | 0.268 | 0.185 | 0.107 | 0.102 | 0.208 |
| 2011 | 0.003 | 0.035 | 0.130 | 0.381 | 0.384 | 0.377 | 0.273 | 0.139 | 0.122 | 0.298 |
| 2012 | 0.002 | 0.032 | 0.124 | 0.166 | 0.243 | 0.234 | 0.165 | 0.094 | 0.084 | 0.178 |
| 2013 | 0.002 | 0.036 | 0.162 | 0.160 | 0.247 | 0.229 | 0.159 | 0.086 | 0.086 | 0.189 |
| 2014 | 0.002 | 0.036 | 0.267 | 0.312 | 0.319 | 0.395 | 0.198 | 0.117 | 0.109 | 0.299 |
| 2015 | 0.004 | 0.040 | 0.412 | 0.453 | 0.355 | 0.503 | 0.367 | 0.166 | 0.146 | 0.407 |
| 2016* | 0.004 | 0.051 | 0.317 | 0.397 | 0.421 | 0.493 | 0.316 | 0.154 | 0.154 | 0.378 |

Table 13.3.4. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Estimates of stock numbers at age (thousands) from the final TSA assessment. Estimates refer to January $1^{\text {st }}$, except for age 0 for estimates refer to July $1^{\text {st. }}$. TSA estimated survivors.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 9334178 | 13726648 | 2123425 | 79500 | 45151 | 396642 | 7157 | 435 | 1167 |
| 1973 | 35028610 | 2087849 | 2828264 | 489139 | 17647 | 11305 | 117984 | 2057 | 456 |
| 1974 | 69453988 | 7834434 | 428189 | 672552 | 122488 | 4968 | 3486 | 34581 | 716 |
| 1975 | 4607792 | 15672055 | 1593991 | 109338 | 192870 | 33908 | 1681 | 1108 | 10685 |
| 1976 | 7190961 | 993133 | 3137669 | 353684 | 28542 | 49142 | 10193 | 461 | 3440 |
| 1977 | 15430704 | 1619883 | 225956 | 832447 | 85267 | 8446 | 13555 | 3196 | 1256 |
| 1978 | 31724047 | 3120387 | 301522 | 65069 | 259274 | 21528 | 2655 | 4539 | 1534 |
| 1979 | 64106952 | 6217948 | 574825 | 78136 | 16580 | 63750 | 5474 | 754 | 1763 |
| 1980 | 11637362 | 12466923 | 1161388 | 141710 | 18194 | 4724 | 18826 | 1693 | 789 |
| 1981 | 19582587 | 2237958 | 2384127 | 343941 | 33952 | 4727 | 1645 | 6253 | 825 |
| 1982 | 11819198 | 3813036 | 439659 | 795156 | 101210 | 10819 | 1789 | 637 | 2590 |
| 1983 | 38209077 | 2324212 | 762113 | 162897 | 296137 | 37801 | 4737 | 802 | 1401 |
| 1984 | 7549963 | 7486807 | 479856 | 262509 | 48580 | 93661 | 12301 | 1830 | 841 |
| 1985 | 12359046 | 1661260 | 1563731 | 161127 | 71547 | 12746 | 30364 | 4346 | 919 |
| 1986 | 23436516 | 2528814 | 366927 | 552250 | 45553 | 20118 | 4283 | 10792 | 1961 |
| 1987 | 511175 | 4506975 | 581486 | 112248 | 151249 | 11673 | 6798 | 1636 | 4756 |
| 1988 | 1442810 | 387952 | 1100841 | 163695 | 29628 | 43666 | 3861 | 2322 | 2353 |
| 1989 | 2623166 | 613286 | 109440 | 365889 | 36268 | 7742 | 13456 | 1347 | 1747 |
| 1990 | 11049117 | 840994 | 154510 | 35488 | 103842 | 9321 | 2616 | 4787 | 1207 |
| 1991 | 12277185 | 2520843 | 218902 | 42619 | 9778 | 30428 | 3229 | 1042 | 2522 |
| 1992 | 20621156 | 2846325 | 639791 | 69374 | 11512 | 2738 | 9664 | 1171 | 1352 |
| 1993 | 5212195 | 4992663 | 758672 | 216064 | 17893 | 3175 | 1058 | 3312 | 1030 |
| 1994 | 20248492 | 1308150 | 1287523 | 220415 | 58795 | 4917 | 970 | 371 | 1618 |
| 1995 | 5600800 | 5262795 | 352600 | 401403 | 58323 | 16679 | 1396 | 304 | 734 |
| 1996 | 7872209 | 1485968 | 1446139 | 127912 | 118251 | 17208 | 5784 | 462 | 432 |
| 1997 | 4641366 | 2116660 | 407633 | 561384 | 39399 | 32835 | 5159 | 1820 | 370 |
| 1998 | 3468782 | 1260358 | 564725 | 165040 | 217714 | 14089 | 10510 | 1913 | 996 |
| 1999 | 52278672 | 953361 | 324020 | 196827 | 60412 | 68813 | 4828 | 3753 | 1317 |
| 2000 | 10186254 | 14101149 | 245511 | 105836 | 55346 | 19108 | 17971 | 1599 | 2163 |
| 2001 | 958698 | 2719783 | 3632203 | 77821 | 28614 | 15542 | 6415 | 5971 | 1732 |
| 2002 | 1341073 | 372385 | 693827 | 1568301 | 27767 | 10464 | 6126 | 2798 | 4154 |
| 2003 | 1508352 | 424676 | 90586 | 341421 | 791523 | 12671 | 5058 | 3192 | 4294 |
| 2004 | 1460362 | 455286 | 105751 | 47188 | 198856 | 455819 | 6998 | 3024 | 5104 |
| 2005 | 14398949 | 453617 | 110862 | 54434 | 26973 | 116413 | 256715 | 4286 | 5653 |
| 2006 | 3033564 | 4076445 | 110164 | 53040 | 28103 | 15483 | 64525 | 147431 | 6806 |
| 2007 | 1992846 | 891430 | 1007711 | 45343 | 23037 | 12369 | 7092 | 34341 | 95103 |
| 2008 | 1374472 | 629101 | 225611 | 492626 | 20040 | 10472 | 5852 | 3846 | 84705 |
| 2009 | 10329091 | 492743 | 167007 | 115138 | 286885 | 10799 | 5881 | 3582 | 62850 |
| 2010 | 952186 | 3379676 | 136053 | 88640 | 69487 | 165175 | 6407 | 3867 | 49088 |
| 2011 | 113917 | 370349 | 956680 | 68779 | 50960 | 41009 | 95146 | 4209 | 39121 |
| 2012 | 1300389 | 129841 | 106505 | 489231 | 33375 | 25425 | 20882 | 57038 | 31377 |
| 2013 | 576257 | 477482 | 37955 | 53689 | 293098 | 18851 | 14655 | 13867 | 65554 |
| 2014 | 5885774 | 268002 | 140311 | 17819 | 32041 | 162858 | 10719 | 9727 | 59499 |
| 2015 | 1488346 | 2108931 | 78915 | 60264 | 8929 | 16608 | 78677 | 6862 | 50690 |
| 2016* | 3279669 | 541246 | 617993 | 29394 | 26822 | 4460 | 7211 | 42575 | 40651 |

Table 13.3.5. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Stock summary table. Both estimates (EST) and standard errors (SE) are given. *TSA model fits or projections.

| $\begin{aligned} & \stackrel{1}{0} \\ & \underset{\sim}{\sim} \end{aligned}$ | $\underset{\text { ָ }}{\substack{\text { ָ } \\ \hline}}$ |  | $\begin{aligned} & \underset{\sim}{n} \\ & \underset{U}{U} \\ & \tilde{U} \end{aligned}$ | $\begin{aligned} & \text { n } \\ & \text { ㅡ } \\ & \text { ㄷ } \\ & \text { त्} \end{aligned}$ |  | $\begin{aligned} & \ddot{\sim} \\ & \dot{y} \\ & \tilde{y} \\ & \underset{\sim}{c} \\ & \underline{\pi} \end{aligned}$ |  | $\begin{aligned} & \tilde{n} \\ & \tilde{n} \\ & \tilde{0} \\ & \tilde{N} \\ & \ddot{0} \end{aligned}$ | $\begin{aligned} & \tilde{n} \\ & n \\ & \tilde{0} \\ & \tilde{\pi} \\ & \tilde{i} \end{aligned}$ |  |  | $\begin{aligned} & \stackrel{\rightharpoonup}{0} \\ & \dot{\omega} \\ & \tilde{\omega} \end{aligned}$ | $\begin{aligned} & \stackrel{\sim}{n} \\ & \tilde{\omega} \end{aligned}$ | $\begin{aligned} & \stackrel{\omega}{\omega} \\ & \dot{\omega} \\ & \stackrel{\omega}{2} \end{aligned}$ | $\begin{aligned} & \stackrel{\sim}{n} \\ & \tilde{\sim} \end{aligned}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 408148 | 384782 | 40729 | 233286 | 229555 | 24740 | 173903 | 155227 | 28552 | 0.852 | 0.064 | 294835 | 29279 | 2648076 | 264743 | 9334178 | 2149032 |
| 1973 | 344574 | 383454 | 52636 | 206808 | 217530 | 21257 | 137198 | 165924 | 41299 | 0.775 | 0.070 | 276958 | 18789 | 2705509 | 250240 | 35028610 | 4382995 |
| 1974 | 397035 | 250556 | 29034 | 167410 | 159239 | 14229 | 229503 | 91318 | 22137 | 0.723 | 0.071 | 331762 | 23871 | 3175524 | 315271 | 69453988 | 9853342 |
| 1975 | 494390 | 300148 | 41267 | 159640 | 161042 | 14059 | 334009 | 139106 | 35223 | 0.848 | 0.081 | 161530 | 11149 | 2295633 | 261002 | 4607792 | 1839161 |
| 1976 | 402046 | 344062 | 52660 | 181758 | 206412 | 23035 | 217725 | 137650 | 40566 | 0.792 | 0.080 | 196822 | 16286 | 1131585 | 129510 | 7190961 | 1732580 |
| 1977 | 240292 | 200014 | 21973 | 155748 | 163026 | 18214 | 83726 | 36989 | 9142 | 0.802 | 0.084 | 353701 | 30047 | 895899 | 72118 | 15430704 | 2137114 |
| 1978 | 146664 | 138616 | 13473 | 102222 | 102560 | 10026 | 43760 | 36056 | 7139 | 0.890 | 0.085 | 159401 | 13687 | 909472 | 58978 | 31724047 | 2439629 |
| 1979 | 149126 | 143546 | 16046 | 97389 | 86482 | 9183 | 51376 | 57064 | 10363 | 0.909 | 0.087 | 93894 | 10621 | 1414100 | 78246 | 64106952 | 5640228 |
| 1980 | 202624 | 189849 | 19295 | 110832 | 106225 | 10345 | 91263 | 83625 | 14195 | 0.878 | 0.080 | 103563 | 10442 | 1573575 | 94418 | 11637362 | 1266833 |
| 1981 | 226556 | 225290 | 21210 | 147673 | 151454 | 14970 | 78665 | 73836 | 12176 | 0.664 | 0.062 | 193503 | 12682 | 1133434 | 58852 | 19582587 | 1993292 |
| 1982 | 256280 | 208783 | 15508 | 195359 | 167041 | 13241 | 60730 | 41742 | 6834 | 0.548 | 0.046 | 432650 | 20208 | 958749 | 39914 | 11819198 | 1042453 |
| 1983 | 253314 | 226786 | 16637 | 187963 | 178615 | 12875 | 64451 | 48170 | 7615 | 0.710 | 0.053 | 294486 | 15147 | 1252502 | 50836 | 38209077 | 2639044 |
| 1984 | 247209 | 227552 | 22634 | 157631 | 150013 | 11103 | 89057 | 77539 | 16956 | 0.832 | 0.060 | 238292 | 14740 | 1522315 | 79209 | 7549963 | 1647606 |
| 1985 | 247401 | 226526 | 18400 | 182552 | 165805 | 13901 | 64375 | 60721 | 9878 | 0.787 | 0.056 | 168059 | 8494 | 1000381 | 43200 | 12359046 | 1491235 |
| 1986 | 223837 | 206451 | 15059 | 184520 | 164171 | 12518 | 38735 | 42280 | 6754 | 0.884 | 0.060 | 288815 | 16382 | 1234667 | 62711 | 23436516 | 2152177 |
| 1987 | 195048 | 178828 | 14848 | 133892 | 125167 | 9518 | 60046 | 53661 | 9367 | 0.890 | 0.062 | 163463 | 8896 | 885459 | 45447 | 511175 | 1352857 |
| 1988 | 179912 | 168320 | 13929 | 124800 | 122018 | 10852 | 53729 | 46303 | 7396 | 0.939 | 0.067 | 126941 | 8558 | 506170 | 95818 | 1442810 | 2089310 |
| 1989 | 127674 | 117973 | 9869 | 91927 | 93147 | 8565 | 34872 | 24827 | 4388 | 0.888 | 0.067 | 178229 | 11191 | 426183 | 60207 | 2623166 | 1530656 |
| 1990 | 86734 | 78217 | 7472 | 61187 | 56924 | 5138 | 25159 | 21293 | 4070 | 0.886 | 0.066 | 85310 | 5813 | 732903 | 67940 | 11049117 | 1587267 |
| 1991 | 97213 | 91860 | 12597 | 54731 | 45418 | 4557 | 41993 | 46442 | 10219 | 0.875 | 0.066 | 52414 | 3953 | 929299 | 46914 | 12277185 | 942659 |
| 1992 | 135092 | 126306 | 11938 | 80479 | 71615 | 7209 | 53421 | 54691 | 8322 | 0.874 | 0.054 | 55814 | 2850 | 933717 | 41636 | 20621156 | 1497217 |
| 1993 | 180233 | 210707 | 21101 | 97867 | 109840 | 10529 | 81509 | 100867 | 16260 | 0.932 | 0.058 | 119014 | 7469 | 934274 | 46750 | 5212195 | 469606 |
| 1994 | 169501 | 233408 | 21969 | 94708 | 129957 | 13396 | 74297 | 103451 | 14983 | 0.911 | 0.060 | 139261 | 9711 | 980038 | 42277 | 20248492 | 1367437 |
| 1995 | 168825 | 176392 | 17266 | 89581 | 104415 | 10806 | 79035 | 71977 | 11667 | 0.820 | 0.059 | 197825 | 14114 | 930407 | 44504 | 5600800 | 447773 |
| 1996 | 204701 | 201451 | 18802 | 92422 | 98385 | 8892 | 112055 | 103066 | 14684 | 0.805 | 0.057 | 124912 | 7260 | 833852 | 36272 | 7872209 | 659512 |
| 1997 | 170055 | 164910 | 14764 | 95341 | 95272 | 8938 | 74603 | 69638 | 10377 | 0.629 | 0.050 | 253112 | 14918 | 793948 | 36658 | 4641366 | 460477 |
| 1998 | 161962 | 159278 | 13704 | 95387 | 92507 | 7704 | 66457 | 66770 | 9456 | 0.735 | 0.057 | 182752 | 9800 | 613248 | 27410 | 3468782 | 326372 |
| 1999 | 123449 | 128048 | 11106 | 75872 | 73363 | 6129 | 47446 | 54685 | 7637 | 0.824 | 0.063 | 141729 | 8884 | 1687959 | 100450 | 52278672 | 3809693 |
| 2000 | 126876 | 169990 | 30586 | 54358 | 55577 | 5125 | 72395 | 114414 | 28199 | 0.893 | 0.067 | 91483 | 6414 | 2390258 | 136384 | 10186254 | 697739 |
| 2001 | 173525 | 279235 | 38737 | 47383 | 100043 | 14749 | 125978 | 179192 | 31339 | 0.614 | 0.054 | 61624 | 4379 | 1186775 | 68898 | 958698 | 749473 |
| 2002 | 155164 | 190651 | 22550 | 64778 | 100630 | 12406 | 89745 | 90021 | 16451 | 0.387 | 0.038 | 523362 | 36523 | 763676 | 41012 | 1341073 | 390034 |
| 2003 | 74412 | 101470 | 11681 | 46991 | 78392 | 9735 | 27149 | 23078 | 4491 | 0.244 | 0.026 | 450852 | 27870 | 541895 | 30070 | 1508352 | 365457 |
| 2004 | 72510 | 78848 | 9545 | 51760 | 67481 | 8680 | 20586 | 11367 | 2029 | 0.243 | 0.026 | 312235 | 22259 | 457643 | 25547 | 1460362 | 228366 |
| 2005 | 64115 | 65826 | 7768 | 51437 | 56352 | 7117 | 12573 | 9474 | 1487 | 0.307 | 0.031 | 234038 | 19910 | 1114204 | 48479 | 14398949 | 763580 |
| 2006 | 66955 | 66908 | 8192 | 43185 | 45981 | 5509 | 23622 | 20927 | 4530 | 0.503 | 0.044 | 163429 | 16391 | 835647 | 35576 | 3033564 | 209238 |
| 2007 | 67437 | 76402 | 8112 | 34572 | 45800 | 5253 | 32751 | 30603 | 4741 | 0.423 | 0.039 | 136082 | 16431 | 590269 | 29061 | 1992846 | 354733 |
| 2008 | 47730 | 56603 | 5846 | 30755 | 41562 | 4420 | 14694 | 15041 | 2623 | 0.250 | 0.026 | 282911 | 19532 | 488849 | 24422 | 1374472 | 308960 |
| 2009 | 47943 | 44874 | 4484 | 34614 | 36609 | 3805 | 12372 | 8266 | 1287 | 0.195 | 0.021 | 236544 | 18807 | 869507 | 34393 | 10329091 | 526309 |
| 2010 | 45411 | 44321 | 4631 | 31460 | 35427 | 3638 | 13473 | 8895 | 1731 | 0.208 | 0.022 | 221500 | 18310 | 539533 | 27284 | 952186 | 747724 |
| 2011 | 49677 | 57960 | 5572 | 36392 | 40482 | 3712 | 13082 | 17479 | 3027 | 0.298 | 0.030 | 158629 | 12012 | 461151 | 21473 | 113917 | 557619 |
| 2012 | 43200 | 45180 | 4530 | 37619 | 39652 | 3971 | 5032 | 5528 | 1133 | 0.178 | 0.020 | 332903 | 18651 | 445947 | 21569 | 1300389 | 248483 |
| 2013 | 47068 | 42278 | 4399 | 43631 | 38898 | 4071 | 3352 | 3380 | 699 | 0.189 | 0.020 | 262506 | 14008 | 380357 | 18312 | 576257 | 235123 |
| 2014 | 46316 | 50618 | 5049 | 39765 | 45473 | 4636 | 5149 | 5145 | 1020 | 0.299 | 0.031 | 184867 | 11635 | 530077 | 30181 | 5885774 | 606801 |
| 2015 | 41596 | 50330 | 5206 | 34376 | 40719 | 3911 | 6285 | 9611 | 2417 | 0.407 | 0.045 | 146053 | 10656 | 530895 | 36692 | 1488346 | 301125 |
| 2016* |  | 77855 | 19990 |  | 62463 | 16889 |  | 15392 | 5431 | 0.378 | 0.098 |  |  |  |  | 3279669 | 896276 |

Table 13.6.1. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Short-term forecast input.

| MFDP version 1a |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RuN: IBP02 |  |  |  |  |  |  |
| Time and date: 12:09 14/10/2016 |  |  |  |  |  |  |
| Fbar age range (Total) : 2-4 |  |  |  |  |  |  |
| Fbar age range Fleet 1: 2-4 |  |  |  |  |  |  |
| Fbar age range Fleet $2: \mathbf{2 - 4}$ |  |  |  |  |  |  |
| 2016 |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt |
| 0 | 3279669 | 1.02 | 0 | 0 | 0 | 0.036 |
| 1 | 541246 | 1.19 | 0 | 0 | 0 | 0.152 |
| 2 | 617993 | 0.58 | 0 | 0 | 0 | 0.425 |
| 3 | 29394 | 0.36 | 1 | 0 | 0 | 0.580 |
| 4 | 26822 | 0.34 | 1 | 0 | 0 | 0.760 |
| 5 | 4460 | 0.34 | 1 | 0 | 0 | 0.958 |
| 6 | 7211 | 0.25 | 1 | 0 | 0 | 0.936 |
| 7 | 42575 | 0.22 | 1 | 0 | 0 | 0.763 |
| 8 | 40651 | 0.2 | 1 | 0 | 0 | 0.948 |
| Catch |  |  |  |  |  |  |
| Age | Sel | CWt | DSel | DCWt |  |  |
| 0 | 0.000 | 0.000 | 0.004 | 0.036 |  |  |
| 1 | 0.002 | 0.356 | 0.048 | 0.141 |  |  |
| 2 | 0.198 | 0.523 | 0.119 | 0.267 |  |  |
| 3 | 0.344 | 0.620 | 0.052 | 0.369 |  |  |
| 4 | 0.405 | 0.654 | 0.016 | 0.479 |  |  |
| 5 | 0.475 | 1.016 | 0.018 | 0.497 |  |  |
| 6 | 0.311 | 0.857 | 0.006 | 0.551 |  |  |
| 7 | 0.153 | 0.708 | 0.001 | 0.491 |  |  |
| 8 | 0.154 | 0.83 | 0.001 | 0.714 |  |  |
| IBC |  |  |  |  |  |  |
| Age | Sel | CWt |  |  |  |  |
| 0 | 0.000 | 0 |  |  |  |  |
| 1 | 0.000 | 0.3562 |  |  |  |  |
| 2 | 0.000 | 0.5228 |  |  |  |  |
| 3 | 0.000 | 0.6205 |  |  |  |  |
| 4 | 0.000 | 0.6425 |  |  |  |  |
| 5 | 0.001 | 0.6619 |  |  |  |  |
| 6 | 0.000 | 0.7576 |  |  |  |  |
| 7 | 0.000 | 0.7728 |  |  |  |  |
| 8 | $0$ | 1.41 |  |  |  |  |

Table 13.6.1 (cont).
Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Short-term forecast input.

2017

| Age | N | M | Mat | PF | PM | SWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 3279669 | 1.020 | 0 | 0 | 0 | 0.036 |
| 1 | $\cdot$ | 1.190 | 0 | 0 | 0 | 0.152 |
| 2 | $\cdot$ | 0.580 | 0 | 0 | 0 | 0.425 |
| 3 | $\cdot$ | 0.360 | 1 | 0 | 0 | 0.583 |
| 4 | $\cdot$ | 0.340 | 1 | 0 | 0 | 0.771 |
| 5 | $\cdot$ | 0.340 | 1 | 0 | 0 | 0.944 |
| 6 | $\cdot$ | 0.250 | 1 | 0 | 0 | 1.148 |
| 7 | $\cdot$ | 0.22 | 1 | 0 | 0 | 1 |
| 8 | $\cdot$ | 0.2 | 1 | 0 | 0 | 1 |


| Catch |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Age | Sel | CWt | DSel | DCWt |
| 0 | 0.000 | 0.000 | 0.004 | 0.036 |
| 1 | 0.002 | 0.356 | 0.048 | 0.141 |
| 2 | 0.198 | 0.523 | 0.119 | 0.267 |
| 3 | 0.344 | 0.620 | 0.052 | 0.348 |
| 4 | 0.405 | 0.642 | 0.016 | 0.482 |
| 5 | 0.475 | 0.726 | 0.018 | 0.587 |
| 6 | 0.311 | 1.197 | 0.006 | 0.588 |
| 7 | 0.153 | 0.957 | 0.001 | 1 |
| 8 | 0.154 | 0.87 | 0.001 | 1 |


| IBC |  |  |
| :--- | :--- | :--- |
| Age | Sel | CWt |
| 0 | 0.000 | 0 |
| 1 | 0.000 | 0.3562 |
| 2 | 0.000 | 0.5228 |
| 3 | 0.000 | 0.6205 |
| 4 | 0.000 | 0.6425 |
| 5 | 0.001 | 0.6619 |
| 6 | 0.000 | 0.7576 |
| 7 | 0 | 0.7728 |
| 8 | 0 | 1.41 |

Table 13.6.1 (cont).
Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Short-term forecast input.

| 2018 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M | Mat | PF | PM | SWt |
| 0 | 3279669 | 1.02 | 0 | 0 | 0 | 0.036 |
| 1 | . | 1.19 | 0 | 0 | 0 | 0.152 |
| 2 | . | 0.58 | 0 | 0 | 0 | 0.425 |
| 3 | . | 0.36 | 1 | 0 | 0 | 0.583 |
| 4 | . | 0.34 | 1 | 0 | 0 | 0.632 |
| 5 | . | 0.34 | 1 | 0 | 0 | 0.963 |
| 6 | . | 0.25 | 1 | 0 | 0 | 1 |
| 7 | . | 0.22 | 1 | 0 | 0 | 1 |
| 8 | . | 0.2 | 1 | 0 | 0 | 1 |
| Catch |  |  |  |  |  |  |
| Age | Sel | CWt | DSel | DCWt |  |  |
| 0 | 0 | 0 | 0.004 | 0.036 |  |  |
| 1 | $0.002$ | 0.356 | 0.048 | 0.141 |  |  |
| 2 | 0.198 | 0.523 | 0.119 | 0.267 |  |  |
| 3 | 0.344 | 0.62 | 0.052 | 0.348 |  |  |
| 4 | 0.405 | 0.642 | 0.016 | 0.372 |  |  |
| 5 | 0.475 | 0.662 | 0.018 | 0.596 |  |  |
| 6 | 0.311 | 0.799 | 0.006 | 0.696 |  |  |
| 7 | 0.153 | 1.379 | 0.001 | 0.679 |  |  |
| 8 | 0.154 | 0.966 | 0.001 | 0.888 |  |  |
| IBC |  |  |  |  |  |  |
| Age | Sel | CWt |  |  |  |  |
| 0 | 0 | 0.000 |  |  |  |  |
| 1 | 0 | 0.356 |  |  |  |  |
| 2 | 0 | 0.523 |  |  |  |  |
| 3 | 0 | 0.621 |  |  |  |  |
| 4 | 0 | 0.643 |  |  |  |  |
| 5 | 0.001 | 0.662 |  |  |  |  |
| 6 | 0 | 0.7576 |  |  |  |  |
| 7 | 0 | 0.7728 |  |  |  |  |
| 8 | 0 | 1.41 |  |  |  |  |

Table 13.6.2. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Short-term forecast output. A number of management options are highlighted.

|  |  |  |  | $\begin{aligned} & \stackrel{*}{*} \\ & \stackrel{\sim}{O} \\ & \underset{N}{U} \\ & \underset{\sim}{n} \end{aligned}$ | $\begin{aligned} & \tilde{n} \\ & \tilde{\sim} \end{aligned}$ |  | $\begin{aligned} & N \\ & \underset{N}{N} \\ & \underset{O}{O} \\ & \underset{\sim}{C} \end{aligned}$ | $\begin{aligned} & N \\ & \underset{\sim}{N} \\ & \underset{y}{n} \\ & \underline{y} \end{aligned}$ | $\begin{aligned} & N \\ & \underset{N}{N} \\ & \underset{O}{O} \\ & \underset{\sim}{4} \end{aligned}$ | $\begin{aligned} & \infty \\ & \underset{\sim}{0} \\ & \underset{\sim}{n} \end{aligned}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MSY | 39.461 | 33.385 | 6.071 | 0.006 | New F(msy) estimate | 0.190 | 0.159 | 0.031 | 0.000 | 205.595 | -13\% | -45\% | 12\% |
| Management plan | 60.003 | 50.653 | 9.344 | 0.006 | MP target F | 0.300 | 0.251 | 0.049 | 0.000 | 187.292 | -20\% | -17\% | 70\% |
| IBC only | 0.007 | 0.000 | 0.000 | 0.007 | No HC fishery | 0.000 | 0.000 | 0.000 | 0.000 | 241.562 | 3\% | -100\% | -100\% |
| Other options | 57.029 | 48.159 | 8.864 | 0.006 | $0.75{ }^{*} \mathrm{~F}(\mathrm{sq})$ | 0.284 | 0.237 | 0.047 | 0.000 | 189.932 | -19\% | -21\% | 61\% |
|  | 73.469 | 61.914 | 11.549 | 0.006 | Fsq | 0.378 | 0.316 | 0.062 | 0.000 | 175.332 | -25\% | 2\% | 108\% |
|  | 88.817 | 74.691 | 14.118 | 0.005 | 1.25 * F(sq) | 0.473 | 0.395 | 0.078 | 0.000 | 161.934 | -31\% | 23\% | 151\% |
|  | 64.940 | 54.789 | 10.145 | 0.006 | 15\% TAC decrease | 0.328 | 0.274 | 0.054 | 0.000 | 182.918 | -22\% | -15\% | 84\% |
|  | 75.903 | 63.949 | 11.947 | 0.006 | Rollover TAC | 0.392 | 0.327 | 0.065 | 0.000 | 173.245 | -26\% | 0\% | 115\% |
|  | 86.290 | 72.594 | 13.689 | 0.005 | 15\% TAC increase | 0.456 | 0.381 | 0.075 | 0.000 | 164.140 | -30\% | 15\% | 144\% |
|  | 55.300 | 46.708 | 8.586 | 0.006 | $\mathrm{F}(\mathrm{pa})$ | 0.274 | 0.229 | 0.045 | 0.000 | 191.468 | -19\% | -24\% | 57\% |
|  | 72.196 | 60.856 | 11.334 | 0.006 | F (msy) long timeseries | 0.370 | 0.309 | 0.061 | 0.000 | 176.509 | -25\% | 0\% | 104\% |
|  | 135.308 | 112.682 | 22.625 | 0.005 | F(pa) long timeseries | 0.830 | 0.693 | 0.137 | 0.000 | 122.389 | -48\% | 87\% | 283\% |
|  | 73.882 | 62.263 | 11.612 | 0.006 | F(lim) | 0.380 | 0.317 | 0.063 | 0.000 | 175.024 | -26\% | 2\% | 109\% |
|  | 146.572 | 121.590 | 24.985 | 0.005 | B(lim) | 0.946 | 0.790 | 0.156 | 0.000 | 94.000 | -52\% | 103\% | 315\% |
|  | 119.463 | 99.898 | 19.561 | 0.005 | B(trigger) | 0.692 | 0.578 | 0.114 | 0.000 | 132.000 | -42\% | 65\% | 238\% |

* "Wanted" and "unwanted" catch are used to described fish that would be landed and discarded in the absence of the EU landing obligation based on discard rates estimates for 2013-2015.
** Industrial bycatch (IBC) also based on average proportion of the total catch for 2013-2015.
${ }^{1}$ SSB 2018 relative to SSB 2017.
${ }^{2}$ Total catch in 2017 relative to the combined TACs 2016 (TAC $4=61.933$; TAC 3.a $=3.926$; TAC 6.a $=6.462$; Total $=72.321$ ).
${ }^{3}$ Total catch in 2017 relative to WG estimates of wanted catch in 2015


Figure 13.2.1: Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Reported landings for each sampled and unsampled fleet in the full stock area, along with cumulative landings for fleets in descending order of yield.


Figure 13.2.2: Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Summary of landings for fleets with and without discard estimates.


Figure 13.2.3. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Yield by catch component.


Figure 13.2.4. Haddock in Subarea 4 and Divisions 3.a. Proportion of total catch discarded, by age and year.


Figure 13.2.5. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Mean weights-at-age (kg) by catch component. Total catch mean weights are also used as stock mean weights. Red dotted lines give loess smoothers through each time-series of mean weights-at-age.


Figure 13.2.6. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Time series of estimated natural mortality at age, from ICES-WGSAM (2014).


Figure 13.2.7. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Survey distributions by age for the international IBTS Q1 survey (North Sea).


Figure 13.2.8. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Survey distributions by age for the international IBTS Q3 survey (North Sea).


Figure 13.2.9. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Survey distributions by age and quarter for the Scottish West Coast Q1 survey (West of Scotland). Rows show years 2013-2016 (from top to bottom).


Figure 13.2.10. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Survey log CPUE (catch per unit effort) at age.


Figure 13.3.1. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Log catch curves by cohort for total catches.


Figure 13.3.2. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Negative gradients of $\log$ catches per cohort, averaged over ages 2-4. The x-axis represents the spawning year of each cohort.


Northern Shelf haddock (IV, IIla, Vla). Commercial catch correlations

Figure 13.3.3. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Correlations in the catch-at-age matrix (including the plus-group for ages 8), comparing estimates at different ages for the same year-classes (cohorts). In each plot, the straight line is a normal linear model fit: a thick line (and black points) represents a significant ( $p<0.05$ ) regression, while a thin line (and blue points) is not significant. Approximate $95 \%$ confidence intervals for each fit are also shown.


Figure 13.3.4. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Summary plots from an exploratory SAM assessment. Time-series of estimated SSB (top left), mean F(2-4) (top right) and recruitment (bottom right) are shown with approximate pointwise $95 \%$ confidence intervals. Retrospective runs are included in these plots. Model residuals (bottom left) are depicted with a clear blue circle for a positive residual, and a solid red circle for a negative residual.


Figure 13.3.5. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Summary plots from an exploratory SURBAR assessment, using both available surveys (IBTS Q1 and Q3). Mean mortality Z (ages 2 to 4 ), relative spawning stock biomass (SSB), relative total biomass (TSB), and relative recruitment. Shaded grey areas correspond to the $90 \% \mathrm{CI}$. Green points give the model estimates, while red crosses and black lines give (respectively) the mean and median values from the uncertainty estimation bootstrap.


Figure 13.3.6. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Log residuals by age from an exploratory SURBAR assessment, using both available surveys (IBTS Q1 and Q3).


Figure 13.3.7. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Log abundance indices by cohort (survey "catch curves") for each of the survey indices.


Figure 13.3.8. Haddock in Subarea 4, Division 3.a20 and Division 6.a. Mean-standardised log abundance indices by age and cohort for each of the survey indices. The age represented by each line is indicated by a circled number at the start of the line.


Figure 13.3.9. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Within-survey correlations for the IBTS Q1 (upper) and Q3 (lower) survey series, comparing index values at different ages for the same year-classes (cohorts). In each plot, the straight line is a normal linear model fit: a thick line (with black points) represents a significant ( $p<0.05$ ) regression, while a thin line (with blue points) is not significant. Approximate $\mathbf{9 5 \%}$ confidence intervals for each fit are also shown.


Figure 13.3.10. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Comparisons of stock summary estimates from TSA (blue), SAM (red) and SURBAR (green) models. To facilitate comparison, values have been mean-standardised using the year range for which estimates are available from all three models, and a composite $Z$ estimate has been made for TSA and SAM by adding natural and fishing mortality estimates.


Figure 13.3.11. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Stock summary from final TSA assessment (including forecasts for 2016). Red lines (or points) give best estimates, grey bands (or lines) give approximate pointwise $95 \%$ confidence intervals, and black points give observed values (for catch, discards+IBC, and landings).


Figure 13.3.12. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Stock-recruitment estimates from the final TSA assessment. Points are labelled by year-class


Figure 13.3.13. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Estimated recruitment time-series from the final TSA assessment. Red points give estimated values with grey bars indicating approximate pointwise $95 \%$ confidence intervals. The black line (also with $95 \% \mathrm{CI}$ ) shows the underlying random-walk recruitment model estimated by TSA.


Figure 13.3.14. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Observed (points) and fitted (red lines with $\mathbf{9 5 \%}$ CI indicated by grey bands) for the proportion discarded by age. Here "discards" is shorthand for combined discards + industrial bycatch. The open points for the years 1973-1977 indicate that these values are treated as missing in the TSA estimation. All haddock of age 0 are assumed to be either discarded or caught as industrial bycatch.


Figure 13.3.15. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Standardised TSA landings prediction errors by age.


Figure 13.3.16. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Standardised TSA discards + IBC prediction errors by age.


Figure 13.3.17. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Standardised TSA prediction errors by age for the IBTS Q1 survey index.


Figure 13.3.18. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Standardised TSA prediction errors by age for the IBTS Q3 survey index.


Figure 13.3.19. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Time-series of observed (points) and fitted (lines) values for total catch, by age.


Figure 13.3.20. Haddock in Subarea 4, Division 3.a.20 and Division 6.a. Time-series of observed (points) and fitted (lines) values for the IBTS Q1 survey index, by age.


Figure 13.3.21. Haddock in Subarea 4, Division 3.a.20 and Division 6.a Time-series of observed (points) and fitted (lines) values for the IBTS Q3 survey index, by age.


Figure 13.3.22. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Retrospective plots for the TSA assessment. The best estimates for each retrospective run end in an open circle, and each run is shown with the approximate pointwise $95 \%$ confidence interval. Estimates and CIs are col-our-coded, with older runs becoming progressively more red.


Figure 13.6.1. Haddock in Subarea 4, Division 3.a.20 and Division 6.a. Results of growth modelling for total catch weights (also used as stock weights) using cohort-based linear models (Jaworski 2011). Cohorts 2008-2013 are shown here. Blue points are available observations, pink dotted lines show linear fits to these points, and pink points indicate projected weights for older ages.


Figure 13.6.2. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Results of growth modelling for landings weights using cohort-based linear models (Jaworski 2011). Cohorts 2008-2012 are shown here. Blue points are available observations, pink dotted lines show linear fits to these points, and pink points indicate projected weights for older ages.


Figure 13.6.3. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Results of growth modelling for discard weights using cohort-based linear models (Jaworski 2011). Cohorts 2008-2013 are shown here. Blue points are available observations, pink dotted lines show linear fits to these points, and pink points indicate projected weights for older ages.


Figure 13.8.1. Haddock in Subarea 4, Division 3.a. 20 and Division 6.a. Results of EqSIM estimation of F (mSY) with the advice error but no rule (top) and of Fp 05 with both advice error and rule (bottom).

## 14 Cod (Gadus morhua) in Subarea 4 and Divisions 7.d and 3.a West (North Sea, Eastern English Channel, Skagerrak)

This assessment relates to the cod stock in the North Sea (Subarea 4), the Skagerrak (the western section of Division 3.a) and the eastern Channel (Division 7.d). This assessment is presented as an update from last year.

A stock annex records more detail and references historic information on the stock definition, ecosystem aspects and the fisheries. This report section records only recent developments and new information presented to WGNSSK. The forecast for this stock was re-run in the Autumn because new information from the IBTS Q3 survey triggered the re-opening criterion - the updated forecast is provided in Annex 04.

### 14.1 General

### 14.1.1 Stock definition

A summary of available information on stock definition can be found in the Stock Annex.

### 14.1.2 Ecosystem aspects

The North Sea is characterised by episodic changes in productivity of key components of the ecosystem. Phytoplankton, zooplankton, demersal and pelagic fish have all exhibited such cycles in variability. Managers should expect long-term change, and ensure that management plans have the potential to respond to new circumstances. Examples of these changes include the gadoid outburst in the 1970s. The contracted range of the North Sea cod stock can be linked to reduced abundance as well as environmental factors. A summary of available information on ecosystem aspects is presented in the Stock Annex.

### 14.1.3 Fisheries

Cod are caught by virtually all the demersal gears in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d, including beam trawls, otter trawls, seine nets, gill nets, trammel nets and lines. Most of these gears take a mixture of species. In some of them, cod are considered to be a bycatch (for example in beam trawls targeting flatfish), and in others the fisheries are directed mainly towards cod (for example, some of the fixed gear fisheries). The main gears landing cod in the EU are primarily TR1 (mainly operated by Scotland, Denmark and Germany), but also GN1 (mainly Denmark), BT1 (mainly Denmark, Belgium and England), BT2 (mainly Netherlands, Belgium and Germany), and TR2. The overall effort by demersal trawls/seines has shown a reduction since 2003, especially in the North Sea. The effort by larger mesh (TR1) had remained relatively stable over the previous cod plan (2004-2009) but has been declining since the full implementation of the new cod plan in 2010 (STECF, 2014). A summary of historic information on the directed and by-catch cod fisheries and past and current technical measures used for the management of cod is presented in the Stock Annex.

## Technical Conservation Measures

In 2009 a new system of effort management, by setting effort ceilings (kilowatt-days), was introduced in accordance with the new cod management plan (EC 1342/2008). The number of kw-days utilized was estimated for the different metiers of the national
fleets during a reference period selected by each nation (2004-2006 or 2005-2007). From these reference values, the effort in the primary métiers catching cod (with discard and bycatch taken into account) would be reduced in direct proportion to reductions in fishing mortality until the new cod management plan target fishing mortality of 0.4 was achieved for levels of SSB at or above $\mathrm{B}_{\mathrm{pa}}$. EC 1342/2008 specifies that the reductions in effort shall be applied to metiers using Otter Trawls, Danish Seines or similar gears with mesh size 80 mm and larger and Gill Nets. However, if certain national fleet segments can provide proof that they use highly selective gears and/or that their catches per fishing trip comprise less than $5 \%$ cod, the reductions will not pertain. National fleet segments with less than $1.5 \%$ cod catches can apply to be excluded from the effort management regime completely. There has been no reduction in effort ceiling in 2013-2016 compared to 2012.

In 2008, Scotland introduced a voluntary programme known as "Conservation Credits", which involved seasonal closures, real-time closures (RTCs) and various selective gear options. This was designed to reduce mortality and discarding of cod. The scheme was incentivised by rewarding participating skippers with additional days at sea. The real-time closures system ( 15 were implemented in 2008) discouraged vessels from operating in areas of high cod abundance. In 2009, the number of closures implemented was increased substantially (to 144 for all areas subject to the cod management plan) and made mandatory, with up to 12 being implemented at any one time. Closures are determined by landings per unit effort, based on fine scale VMS data and daily logbook records and also by on-board inspections. Based on new in-year information on cod movement from tagging, the dimensions of the RTCs were increased by just over four times (from 50 square nautical miles to 225) from July 2010. The use of more species and size selective gears (some trialled by the Marine Laboratory in Aberdeen) formed a further series of options within the scheme. These included the 'Orkney' trawl, the use of nets with 130 mm codends and larger meshes in the square meshed panels of Nephrops trawls. The scheme has delivered a total of 165, 185, 173, 166, 94 and 97 closures in 2010, 2011, 2012, 20132014 and 2015 respectively. ICES notes that from the initial year of operation (2008) cod discarding rates in Scotland have decreased from $61 \%$ to $24 \%$ in 2012, but have increased again to $31 \%$ in $2013,27 \%$ in 2014 and $34 \%$ in 2015; it is hypothesised that this recent increase may be due in part to FDF (fully documented fisheries) vessels putting upward pressure on the lease price of cod, resulting in non-FDF vessels increasing the amount of cod they discard because they are unwilling to pay an above-market price for cod quota.

The expansion of the closed-circuit TV (CCTV) and FDF programmes in 2010-2015 in Scotland, Denmark, Germany, England and the Netherlands is expected to have contributed to the reduction of cod mortality. Under this scheme, UK vessels are not permitted to discard any cod, while Danish, German and Dutch vessels are still permitted to discard undersized cod. For participating vessels, all cod caught are counted against the quota, and in return fishers are permitted additional catches of cod. No effect of changed fishing behaviour has been observed for small vessels ( $<221 \mathrm{kWh}$ ), though changes in fishing behaviour and increased marketable landings have been observed for the larger vessels. Though analyses are still underway, no high grading issues have been detected as yet. Landings by FDF métiers comprised less than $2 \%$ of total landings in 2009, rising to $27 \%$ in 2012, but has since declined to $22 \%$ in 2013 and $21 \%$ in 2014 and 2015 (Intercatch data).

Changes in national fleet dynamics

The ICES WGFTFB report now only provides a description of changes in EU fishing fleets and effort relevant to assessment working groups every second year; there is no such information in recent ICES-WGFTFB $(2014,2015)$ reports.

## The Fishers' North Sea Stock Survey

A fishers' North Sea stock survey for 2015 was not available at the time of the Working Group. Historic comparisons between the fishers' North Sea stock survey and the IBTS survey data are given in previous WGNSSK reports.

### 14.1.4 Management

Management of cod is by TAC and technical measures. The agreed TACs for Cod in Division 3.a (Skagerrak), 7.d and Subarea 4 were as follows:

| TAC(000T) | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3.a <br> (Skagerrak) | 4.1 | 4.8 | 3.8 | 3.8 | 3.8 | 4.0 | 4.2 | 4.8 |
| 2.a +4 | 28.8 | 33.6 | 26.8 | 26.5 | 26.5 | 27.8 | 29.2 | 33.7 |
| 7.d | 1.7 | 2.0 | 1.6 | 1.5 | 1.5 | 1.6 | 1.7 | 2.0 |

For 2009 Council Regulation (EC) ${ }^{\circ} 43 / 2009$ allocates different amounts of $\mathrm{Kw}^{*}$ days by Member State and area to different effort groups of vessels depending on gear and mesh size (see section 2.1.2 for more details). For 2010-2016, Council Regulations (EC) $\mathrm{N}^{\circ} 219 / 2010, \quad \mathrm{~N}^{\circ} 57 / 2011, \quad \mathrm{~N}^{\circ} 44 / 2012$, $\mathrm{N}^{\circ} 297 / 2013$, $\mathrm{N}^{\circ} 432 / 2014, \quad \mathrm{~N}^{\circ} 2015 / 104$ and $\mathrm{N}^{\circ} 2016 / 72$ respectively have updated Council Regulation (EC) $\mathrm{N}^{\circ} 43 / 2009$ with new allocates, based on the same effort groups of vessels and areas as stipulated in Council Regulation (EC) $\mathrm{N}^{\circ} 43 / 2009$.

Demersal fisheries in the area are mixed fisheries, with many stocks exploited together in various combinations in the various fisheries. In these cases, management advice must consider both the state of individual stocks and their simultaneous exploitation in demersal fisheries. Stocks in the poorest condition, particularly those which suffer from reduced reproductive capacity, become the overriding concern for the management of mixed fisheries, where these stocks are exploited either as a targeted species or as a bycatch.

## Cod recovery and management plans

A Cod Recovery Plan which detailed the process of setting TACs for the North Sea cod was in place until 2008. Details of it are given in EC 423/2004 and previous working group reports. ICES considered the recovery plan as not consistent with the precautionary approach because it did not result in a closure of the fisheries for cod at a time of very low stock abundance and until an initial recovery of the cod SSB had been proven.

In April 2008, the European Commission adopted a proposal to amend the cod recovery plan, based on input from stakeholders, and on scientific advice from both ICES and STECF that current measures have been inadequate to reduce fishing pressure on cod to enable stock recovery. The main changes proposed were replacing targets in terms of biomass levels with new targets expressed as optimum fishing rates intended to provide high sustainable yield, and introducing a new system of effort management by setting effort ceilings (kilowatt-days) for groups of vessels or fleet segments to be managed at a national level by Member States. The new system was intended to be
simpler, more flexible and more efficient than the previous one, allowing effort reductions to be proportionate to targeted reductions in fishing mortality for the segments that contribute the most to cod mortality, while for other segments effort will be frozen at the average level for either 2004-2006 or 2005-2007.

In December 2008 the European Commission and Norway agreed on a new cod management plan that aimed to be consistent with the precautionary approach and was intended to achieve sustainable fisheries and high yield, leading to a target fishing mortality of 0.4. In addition to the EU-Norway agreement, the EU has implemented effort restrictions, reducing KW-days available to EU vessels in the main métiers catching cod in direct proportion to reductions in fishing mortality until the long-term phase of the plan is reached, for which the target $F$ is 0.4 if SSB is above $B_{\text {pa. }}$. Details of European Commission plan are given in EC 1342/2008.

A joint ICES-STECF group met during 2011 to conduct a historical evaluation of the effectiveness of these plans (ICES-WKROUNDMP, 2011; Kraak et al., 2013), and concluded that for North Sea cod, although there had been a gradual reduction in F and discards, the plans had not controlled F as envisaged, and that following the current regime was unlikely to deliver $\mathrm{F}_{\text {msy }}$ by 2015. However, there have been positive contributions under Article 13c of the EC plan towards achieving the cod plan targets. The EC plan is currently subject to deliberation between the European Commission, Council and Parliament.

The management plan HCR for setting TAC for North Sea cod stock are as follows (extracts from EC 1342/2008):

Article 7: Procedure for setting TACs for cod stocks in the Kattegat the west of Scotland and the Irish Sea

1. Each year, the Council shall decide on the TAC for the following year for each of the cod stocks in the Kattegat, the west of Scotland and the Irish Sea. The TAC shall be calculated by deducting the following quantities from the total removals of cod that are forecast by STECF as corresponding to the fishing mortality rates referred to in paragraphs 2 and 3:
(a) a quantity of fish equivalent to the expected discards of cod from the stock concerned;
(b) as appropriate a quantity corresponding to other sources of cod mortality caused by fishing to be fixed on the basis of a proposal from the Commission.
2. [assumed to apply to North Sea cod as well] When giving its advice in accordance with paragraphs 2 and 3, STECF shall assume that in the year prior to the year of application of the TAC the stock is fished with an adjustment in fishing mortality equal to the reduction in maximum allowable fishing effort that applies in that year.

Article 8: Procedure for setting TACs for the cod stock in the North Sea, the Skagerrak and the eastern Channel

1. Each year, the Council shall decide on the TACs for the cod stock in the North Sea, the Skagerrak and the eastern Channel. The TACs shall be calculated by applying the reduction rules set out in Article 7 paragraph 1(a) and (b).
2. The TACs shall initially be calculated in accordance with paragraphs 3 and 5. From the year where the TACs resulting from the application of paragraphs 3 and 5 would be lower than the TACs resulting from the application of paragraphs 4 and 5, the TACs shall be calculated according to the paragraphs 4 and 5.
3. Initially, the TACs shall not exceed a level corresponding to a fishing mortality which is a fraction of the estimate of fishing mortality on appropriate age groups in 2008 as follows: $75 \%$ for the TACs in 2009, $65 \%$ for the TACs in 2010, and applying successive decrements of $10 \%$ for the following years.
4. Subsequently, if the size of the stock on 1 January of the year prior to the year of application of the TACs is:
(a) above the precautionary spawning biomass level, the TACs shall correspond to a fishing mortality rate of 0,4 on appropriate age groups;
(b) between the minimum spawning biomass level and the precautionary spawning biomass level, the TACs shall not exceed a level corresponding to a fishing mortality rate on appropriate age groups equal to the following formula: $0,4-(0,2$ * (Precautionary spawning biomass level - spawning biomass) / (Precautionary spawning biomass level - minimum spawning biomass level))
(c) at or below the limit spawning biomass level, the TACs shall not exceed a level corresponding to a fishing mortality rate of 0,2 on appropriate age groups.
5. Notwithstanding paragraphs 3 and 4, the Council shall not set the TACs for 2010 and subsequent years at a level that is more than $20 \%$ below or above the TACs established in the previous year.
6. Where the cod stock referred to in paragraph 1 has been exploited at a fishing mortality rate close to 0,4 during three successive years, the Commission shall evaluate the application of this Article and, where appropriate, propose relevant measures to amend it in order to ensure exploitation at maximum sustainable yield.

Article 9: Procedure for setting TACs in poor data conditions
Where, due to lack of sufficiently accurate and representative information, STECF is not able to give advice allowing the Council to set the TACs in accordance with Articles 7 or 8, the Council shall decide as follows:
(a) where STECF advises that the catches of cod should be reduced to the lowest possible level, the TACs shall be set according to a $25 \%$ reduction compared to the TAC in the previous year;
(b) in all other cases the TACs shall be set according to a $15 \%$ reduction compared to the TAC in the previous year, unless STECF advises that this is not appropriate.

## Article 10: Adaptation of measures

1. When the target fishing mortality rate in Article 5(2) has been reached or in the event that STECF advises that this target, or the minimum and precautionary spawning biomass levels in Article 6 or the levels of fishing mortality rates given in Article 7(2) are no longer appropriate in order to maintain a low risk of stock depletion and a maximum sustainable yield, the Council shall decide on new values for these levels.
2. In the event that STECF advises that any of the cod stocks is failing to recover properly, the Council shall take a decision which:
(a) sets the TAC for the relevant stock at a level lower than that provided for in Articles 7, 8 and 9;
(b) sets the maximum allowable fishing effort at a level lower than that provided for in Article 12;
(c) establishes associated conditions as appropriate.

Changes to the stock assessment and reference points in 2015 imply a need to re-evaluate the management plans in order to ascertain if they can still be considered precautionary under the new stock perception. Until such an evaluation is conducted, advice is given according to the ICES MSY approach.

### 14.2 Data available

### 14.2.1 Catch

Landings data from human consumption fisheries for recent years as officially reported to ICES together with those estimated by the WG are given for each area separately and combined in Table 14.1.

The landings estimate for 2015 is 37.2 thousand tonnes, split as follows for the separate areas (thousand tonnes):

|  |  | TAC | LANDINGS | DISCARDS |
| :--- | :--- | :--- | :--- | :--- |
| 3.a-Skagerrak | 4.2 |  | 4.6 | 2.9 |
| 4 | 29.2 | 31.2 | 9.7 |  |
| $7 . d$ | 1.7 | 1.4 | 0.02 |  |
| Total | 35.1 | 37.2 | 12.6 |  |

WG estimates of discards are also shown in the above table.
Prior to the use of Intercatch for discard estimation, discard numbers-at-age were estimated for areas 4 and 7.d by applying the Scottish discard ogives to the international landings-at-age, and were based on observer sampling estimates for area 3.a-Skagerrak. Discard raising for 2002-2015 was performed in Intercatch, with the different nations providing information by area, quarter and métier. Prior to the reform of the EU's data collection framework in 2008 (see http://datacollection.jrc.ec.europa.eu/), sampling for discards and age compositions was poor in area 7.d, and this necessitated combining areas 4 and $7 . \mathrm{d}$ for 2002 -2008 in order to facilitate computations in Intercatch. The provision of discard information has vastly improved since 2009 and covered $70 \%$ of the landings by weight in 2015, with all nations (apart from Norway) now providing discard information. Figure 14.1a plots reported landings and estimated discards used in the assessment. Discard ratio sampling coverage by area and season for 2015 is provided in Table 14.2e, along with the contributions to total landings and discards from each area prior to raising.

Norwegian discarding is illegal, so although this nation has accounted for $7-14 \%$ of cod landings over the period 2002-2015 (Intercatch data), it does not provide discard estimates. Nevertheless, the agreed procedure applied in Intercatch is that discards raising should include Norway (i.e. Norway will be allocated discards associated with landings in reported métiers). Furthermore, tagging and genetic studies have indicated that Norwegian coastal cod are different to North Sea cod and do not generally move into areas occupied by North Sea cod. Therefore, Norwegian coastal cod data have been removed from North Sea cod data by uploading only North Sea cod data into Intercatch for 2002 onwards, and by adjusting catches prior to 2002 to reflect the removal of Norwegian coastal cod data (an annual multiplicative adjustment of no more
than $2.5 \%$ was made using Norwegian coastal cod data - see ICES-WKNSEA 2015 for more details).

For cod in 4, 3.a-Skagerrak and 7.d, ICES first raised concerns about the mis-reporting and non-reporting of landings in the early 1990s, particularly when TACs became intentionally restrictive for management purposes. Some WG members have since provided estimates of under-reporting of landings to the WG, but by their very nature these are difficult to quantify. In terms of events since the mid-1990s, the WG believes that under-reporting of landings may have been significant in 1998 because of the abundance in the population of the relatively strong 1996 year-class as 2 -year-olds. The landed weight and input numbers at age data for 1998 were adjusted to include an estimated 3000 t of under-reported catch. The 1998 catch estimates remain unchanged in the present assessment (apart from the adjustment for Norwegian coastal cod).

For 1999 and 2000, the WG has no a priori reason to believe that there was significant under-reporting of landings. However, the substantial reduction in fishing effort implied by the 2001, 2002 and 2003 TACs is likely to have resulted in an increase in unreported catch in those years. Anecdotal information from the fisheries in some countries indicated that this may indeed have been the case, but the extent of the alleged underreporting of catch varies considerably.

Marine Scotland-Compliance, a department in the Scottish government responsible for monitoring the Scottish fishing industry, operated a system intended to detect unreported or otherwise illegal fish landings (known as "blackfish"). Records show that blackfish landings have declined significantly since 2003, and is likely to be extremely low since 2006 (ICES-WKCOD, 2011). While the UK Registration of Buyers and Sellers regulation, introduced towards the end of 2005, may have had an important impact on the declining levels of blackfish landings, it is unlikely to be solely responsible, with other factors including large-scale decommissioning, and the development of targeting and monitoring systems that has substantially increased the pressure on the fleet.

The Danish Fisheries Directorate expressed the view that there is no indication of a lack of reporting of cod of any significance for vessels of ten meters and more. This view is based both on the analysis of six indicators of missing reports of landed cod, and a calculation of the difference between the total quantity of cod registered in logbooks and cod registered in sales receipts for Danish vessels over ten meters per quarter over the period 2008-2010, which has been shown to vary between approx. $0.5 \%$ and $2.5 \%$ (ICES-WKCOD, 2011).

Since the WG has no basis to judge the overall extent of under-reported catch over time, it has no alternative but to use its best estimates of landings, which in general are in line with the officially reported landings. An attempt is made to incorporate a catch multiplier to the sum of reported landings and discards data in the assessment of this stock for the period 1993-2005, but the figures shown in Table 14.2c and Figure 14.1a nevertheless comprise the input values to the assessment.

## Age compositions

Age compositions were provided by all nations in 2015, althougth there are gaps from some nations in the years in 2002-2014 (e.g. France prior to 2009, Norway in 2011 and prior to 2005 and the Netherlands prior to 2015). The sampling coverage for landings and discards age compositions for 2015 are reported in Table 14.2e.

Landings in numbers at age for age groups 1-11+ and 1963-2015 are given in Table 14.2a. These data form the basis for the catch at age analysis but do not include industrial fishery by-catches landed for reduction purposes prior to 2002 (values from 2002 onwards were entered into Intercatch for all relevant nations except Norway, and were included in the raising, although the numbers were very small). By-catch estimates are available for the total Danish small-meshed fishery in Division 3.a and Subarea 4 and separately for the Skagerrak (Table 14.1). During the last five years, an average of $74 \%$ of the international landings in number were accounted for by juvenile cod aged 1-3; this average rises to $87 \%$ when considering landings and discards combined. In 2015, age 1 cod comprised $16 \%$ of the total catch by number, age $2,50 \%$ and age $3,22 \%$.

Discard numbers-at-age are shown in Table 14.2b. The proportions of the estimated numbers discarded for ages 1-4 are plotted in Figure 14.1b. The proportion of the estimated total discards by weight are shown in Figure 14.1c, and by number in Figure 14.1d. Estimated proportion of total numbers caught that were discarded (Figure 14.1d) has varied between 35 and $70 \%$ from 1995 to 2005, but has shown an increase to between 70 and $85 \%$ in 2006-8, due to the stronger 2005 year class entering the fishery (estimated to be almost the size of the 1999 year class), and a mismatch between the TAC and effort. The total numbers discarded has decreased to between 50 and $60 \%$ in 2012-15. Historically, the proportion of numbers discarded at age 1 has fluctuated around $80 \%$ with no decline apparent after the introduction of the 120 mm mesh in 2002. Since 2003, it has been at or above $90 \%$, except for a brief decrease to $78 \%$ in 2011 and again in 2014, rising to $86 \%$ in 2015. At ages 2 to 4 discard proportions increased to a maximum around 2006-10, but have subsequently declined to give $58 \%$ for age 2 , $29 \%$ for age 3 and $8 \%$ for 4 year old cod in 2015. Note that these observations refer to numbers discarded, not weight.
Total catch numbers-at-age are shown in Table 14.2c. Landings, discards and total catch numbers at age are given by season in Table 14.2d for 2015. Reported landings, estimated discards and total catch (sum of landings and discards), given in tonnage, are shown in Table 14.4.

## Intercatch

Intercatch was used for estimation of landings, discards and total catch at age and mean weight at age in 2015, and updates performed for 2014 (due to data revisions by UK-England and Wales). Data co-ordinators from each nation were tasked to input data into Intercatch, disaggregated to quarter and métier. The data from Norway excluded Norwegian coastal cod. Allocations of discard ratios and age compositions for unsampled strata were then performed in order to obtain the data required for the assessment. This is the fifth year that Intercatch is used for this purpose for North Sea cod. The approach used for discard ratio allocations was to do it by area (3.a.20, 4 and 7.d) and treat FDF métiers separately, giving six broad categories. Annual discards were first matched to quarterly landings. Then, within each of these six categories, ignoring country and season, where métiers had some samples these were pooled and allocated to unsampled records within that métier. At the end of this process, any remaining métiers were allocated an all-samples pooled discard ratio for the given category.

The landings and discards imported or raised for 2015 are as follows (thousand tonnes; note any differences in landings and discards values to those given above are due to SOP correction):

| CATCH CATEGORY | Raised or Imported |  | CATON |
| :--- | :--- | :--- | :--- |
| PerCentage |  |  |  |
| Discards | Raised | 2.7 | 22 |
| Discards | Imported | 9.9 | 78 |
| Landings | Imported | 37.4 | 100 |

A similar approach was used for allocating age compositions, except that there were 12 broad categories because discards were treated separately to landings.
The landings and discards imported or raised, with age distribution sampled or estimated for 2015 are as follows (thousand tonnes; note any differences in landings and discards values to those given above are due to SOP correction):

| Catch Category | Raised or Imported | Sampled or Estimated | CATON | Percentage |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Landings | Imported | Sampled | 30.3 | 81 |
| Landings | Imported | Estimated | 7.1 | 19 |
| Discards | Imported | Sampled | 9.5 | 75 |
| Discards | Raised | Estimated | 2.7 | 22 |
| Discards | Imported | Estimated | 0.4 | 3 |

Intercatch is discussed in section 1.2, and all results are available on the WGNSSK sharepoint. Further work is ongoing, analysing the Intercatch data (cf ICES WGMIXFISH meeting during 2016).

### 14.2.2 Weight at age

Mean weight at age data for landings, discards and catch, are given in Tables 14.3a-c. Landings, discards and catch mean weights at age are given by season in Table 14.3d for 2015. Total catch mean weight values were also used as stock mean weights. Longterm trends in mean catch weight at age for ages 1-9 are plotted in Figure 14.2, which indicates that there have been short-term trends in mean weight at age and that the decline noted during the 90 's at ages 3 and above now seems to have been reversed. Ages 1 and 2 show little absolute variation over the long-term.

### 14.2.3 Maturity and natural mortality

Until 2015 the maturity values applied to all years were left unchanged from year to year, and were based on NS-IBTS Q1 data from 1981-1985. However, ICES-WKNSEA (2015) noted a change in maturity-at-age in the North Sea cod stock, with fish maturing at a younger age and smaller size. In order to address these changes in the stock, an area-weighted maturity age key was constructed from NS-IBTS Q1 data. As variation in sampling intensity added to the inter-annual variation, a smoother was applied to the maturity age key. This smoothed maturity age key was then applied to the estimation of spawning stock biomass. Maturity in 2016 was based on very low sample sizes ( 8 fish sampled for age 3 in the south), and the WG therefore rejected these maturity estimates and instead smoothed maturities to 2015 and assumed the 2015 maturity values to estimate SSB in 2016. The time-varying maturity ogive used in the assessment is given in Table 14.5a, and the ogives smoothed to either 2015 or 2016 are illustrated in Figure 14.2b.
Table 14.5b and Figure 14.2c show estimates of M, based on multi species considerations adopted for the assessment. ICES-WKROUND (2009) noted that as new stomach data (e.g. on seal predation) become available, a revision of more recent M2 values to reflect the current status of the food web, should be considered. Estimates of natural
mortality, derived from multi-species analyses, are updated by the Working Group on Multi Species Stock Assessment Methods (WGSAM) every three years in so called "key runs" to account for improved knowledge of predation on cod by other species (mainly seals, harbour porpoises and gurnards) and cannibalism; the last update occurred in 2014 with the new key run (ICES-WGSAM 2014). The values presented in Table 14.5b are different to the ones presented by ICES-WGSAM (2014) and ICES-WKNSEA (2015) because an error in the input data to the multi species model was found after WKNSEA that led to a shift in parameters influencing the estimated natural mortalities for cod to a small extent. Between ICES-WGSAM (2014) and ICES-WKNSEA (2015) already unrealistically high predation mortalities on age 3 cod caused by harbour porpoise were corrected in an updated keyrun (see ICES-WKNSEA 2015).

### 14.2.4 Catch, effort and research vessel data

Reliable, individual, disaggregated trip data were not available for the analysis of CPUE. Since the mid-to-late 1990s, changes to the method of recording data means that individual trip data are now more accessible than before; however, the recording of fishing effort as hours fished has become less reliable as it is not a mandatory field in the logbook data. Consequently, the effort data, as hours fished, are not considered to be representative of the fishing effort actually deployed. The WG has previously argued that, although they are in general agreement with the survey information, commercial CPUE tuning series should not be used for the calibration of assessment models due to potential problems with effort recording and hyper-stability (ICES-WGNSSK 2001), and also changes in gear design and usage, as discussed by ICES-WGFTFB (2006, 2007). Therefore, although the commercial fleet series are available, only survey and combined commercial landings and discard information are analysed within the assessment presented.

ICES-WKCOD (2011) analysed UK commercial landings per unit of effort (days fishing) to the northeast and west of Shetland compared to the south and east. Analyses were conducted by gear type and vessel length. Landings per unit of effort (lpue) do not contain discard information or allow for reductions in catch/landings rates resulting from changes in fisher behaviour as part of the Scottish Conservation Credits programme; recent values are therefore likely to be underestimates of the catches and potential catch rates. Vessels from $19-23 \mathrm{~m}$ had a slightly greater increase in their catch rates to the north and west of Shetland, by a factor of 4 compared to 3.5 in the east. When catch rates were averaged across other vessel lengths and across all vessels, the WKCOD analysis could not identify differing rates of increase to either side of the Shetlands but did demonstrate that all vessels have had strong increases in lpue around the Shetlands in recent years.

Two survey series are available for use within this assessment:
Quarter 1 international bottom-trawl survey (IBTS Q1): ages 1-6+, covering the period 1976-2016. This multi-vessel survey covers the whole of the North Sea using fixed stations of at least two tows per rectangle with the GOV trawl.

Quarter 3 international bottom-trawl survey (IBTS Q3): ages 0-6+, covering the period 1991-2015. This multi-vessel survey covers the whole of the North Sea using fixed stations of at least two tows per rectangle with the GOV trawl.

Maps showing the IBTS distribution of cod are presented in Figures 14.3a-b (ages 1$3+$ ). The recent dominant effect of the size and distribution of the 1996 and, to a lesser extent, the 1999, 2005, 2009 and 2013 year-classes are clearly apparent from these charts. Fish of older ages continued to decline until 2006 due to the very weak 2000, 2002 and

2004 year classes, but have subsequently begun to increase, especially in the north and west. The abundance of $3+$ fish is still at a low level compared to historic levels but is increasing. The 2013 year class seems to be distributed more widely when compared to other year classes at the same age, indicating a slightly stronger year class, while the 2014 year class appears to be weak (Figure 14.3a and b), and the 2015 year class even weaker, based on only one survey (the 2016 IBTS Q1 survey; Figure 14.3a; Figure 14.6).

The 2011 benchmark of North Sea Cod resulted in the exclusion of the IBTS Q3 survey index, because divergent trends in recent years were observed when the Q3 index was applied independently of the Q1 index (ICES-WKCOD 2011). At that time it was decided that until the reasons for the discrepancies were resolved, the Q1 was more likely to reflect the stock, and hence the Q3 index was dropped from the assessment. The indices were calculated using the standard stratified mean methodology (mean by rectangle within year, followed by mean over rectangles by year), applied to an extended area (referred to below as the NS-IBTS extended index; ICES-WKROUND 2009; Figure 14.3c). This simple design-based estimator is unable to account for systematic changes in experimental conditions (e.g. change of survey gear). Given these issues, an alternative methodology that calculates standardized age-based survey indices based on GAMs and Delta-distributions (see also Berg WD3, ICES-WKNSEA 2015) has now been adopted (referred to as the NS-IBTS Delta-GAM index), and has led to both the Q1 and Q3 indices being incorporated into the assessment. The general methodology is described in Berg and Kristensen (2012) and Berg et al. (2014) and is implemented in R based on the DATRAS package (http://rforge.net/DATRAS/).

More details of the method used to produce the NS-IBTS Delta-GAM index is provided in the stock annex and can be found in ICES-WKNSEA (2015), as well as the abovementioned publications. In summary the final Delta-GAM models selected for NS-IBTS Q1 and Q3 comprised a stationary spatial model, and included ship, year, depth, time-of-day and haul-duration effects. In addition, the Q3 model also included a gear effect (Q1 only has a single gear, GOV, so this effect is not an issue). The NS-IBTS Delta-GAM indices used in the assessment are given in Table 14.6. Figure 14.3d compares the Q1 and Q3 NS-IBTS extended indices to the corresponding NS-IBTS Delta-GAM indices.

### 14.3 Data analyses

### 14.3.1 Assessment audit

The assessment audit for North Sea cod was completed and no significant issues found. Additional checks on the forecast are carried out during the ICES WGMIXFISH meeting in 2016.

### 14.3.2 Exploratory survey-based analyses

Survey abundance indices are plotted in log-mean standardised form by year and cohort in Figure 14.4a for the IBTS Q1 survey, together with log-abundance curves and associated negative gradients for the age range $2-4$. Similar plots are shown for the IBTS Q3 survey in Figure 14.4b. The log-mean standardised curves indicate no obvious year effects (top-left plots), and tracks cohort signals well (top right). The log abundance curves for each survey series indicate consistent gradients (bottom left), with less steep gradients in recent years (bottom right).

Figures 14.5 a and b show within-survey consistency (in cohort strength) for the NSIBTS Q1 and Q3 Delta-GAM surveys indices, while Figure 14.5c shows between-survey consistency (for each age) for the two surveys. These show generally good consistency,
justifying their use for survey tuning. Correlations deteriorate for age 5 for the IBTS Q3 survey.

The SURBAR survey analysis model was fitted to both the Q1 and Q3 NS-IBTS DeltaGAM survey indices. The summary plots are presented in Figure 14.6.

Biomass -Spawning stock biomass reached the lowest level in the time series in 2005 caused by a series of poor recruitments coupled with high fishing mortality and discard rates at the youngest ages, but SSB has subsequently increased again because of the stronger 2005, 2009 and 2013 year classes and recent reductions in fishing mortality. This increase can also be seen in the time series for total stock biomass.

Total mortality -The SURBAR analysis indicates an overall gradual decline in total mortality with a slight increase in the most recent years.

Recruitment -The SURBAR analysis indicates that the recruiting year classes since 1996 have been relatively weak, but that the 2005, 2009 and 2013 year classes are among the highest of the recent low values.

### 14.3.3 Exploratory catch-at-age-based analyses

## Catch-at-age matrix

The total catch-at-age matrix (Table 14.2c) is expressed as numbers at age, and propor-tions-at-age, standardised over time in Figure 14.7. It shows clearly the contribution of the 1996, 1999, 2005 and 2009 year classes to catches in recent years, with the larger 1996 year class disappearing more rapidly from the catches compared to the 1999, 2005 and 2009 year classes. It also shows the greater proportion of older fish in the catches at the start of the time series relative to recent years, but with the most recent years indicating a relative increase in the number of older fish in the catches. The 2005 and 2009 year classes feature strongly in the catch in the most recent period.

## Catch curve cohort trends

The top panel of Figure 14.8 presents the log catch curve plot for the catch at age data. Through time there is an increase in the slope of the cohort plots indicating faster removal rates or high total mortality. In the most recent years there has been a gradual decrease in the slope at the youngest ages - a sign of decreased mortality rates. The bottom panel plots the negative slope of a regression fitted to the ages $2-4$, the age range used as the reference for mortality trends. The decrease in the negative slope indicates that total mortality rates at the ages comprising the dominant ages within the fishery are declining, with the last three values being the lowest in the time series.

## Assessment model

## SAM

SAM (State-space Assessment Model, Nielsen and Berg 2014) has been used as the assessment model for North Sea cod since 2011, following acceptance at the 2011 benchmark meeting held for the stock (ICES-WKCOD 2011, ICES-WGNSSK 2011). More details can be found in Nielsen and Berg (2014) and in the ICES-WKCOD 2011 report, but essentially SAM models recruitment from a stock-recruitment relationship, with random variability estimated around it, or as a random walk in log-space. Starting from recruitment, each cohort's abundance decreases over time following the usual exponential equation involving natural and fishing mortality. Instead of assuming catches to be known without error and simply subtracting those, SAM assumes that catches
include observation noise, and that the survival process along cohorts is a random process. This has the consequence that estimated F-at-age paths display less interannual variability with SAM than with deterministic assessment models, because part of the observed fluctuations in catch-at-age are arising from observation noise instead of from changes in F .

SAM puts random distributions on the fishing mortalities $F(y, a)$, where ( $y, a)$ denotes year and age. SAM considers a random walk over time for $\log [F(y, a)]$, for each age, allowing for correlation in the increments of the different ages. It has observation equations for both survey indices-at-age and observed catch-at-age, so catch-at-age data are never considered to be known without error. Additionally, in order to deal with the uncertain overall catch levels over the period 1993-2005, SAM estimates annual catch multipliers for this period.

An extension to allow for varying correlation between different ages is achieved by setting the correlation of the $\log \mathrm{F}$ annual increments to be a simple function of the age difference (AR(1) process over the ages). By doing this, individual log F processes will develop correlated in time, but in such a way that neighbouring age classes have more similar fishing mortalities than more distant ones. This correlation structure does not introduce additional parameters to the model, and is referred to as an AR correlation structure (see Nielsen and Berg 2014 for more details).

SAM is considered more appropriate than VPA approaches such as B-Adapt, because the additional variability/uncertainty considered in various components of SAM seems realistic and gives rise to results that are less reactive to noise in the catch or survey data or to potential changes in survey catchability.The fact that SAM considers random variability of the annual survival process along cohorts separately from fishing mortality produces smoother estimated F paths over time. Because the current management regime for the North Sea cod stock is strongly focused on F estimates in the final assessment year, it is important that these estimates do not change too suddenly in response to some data values which may represent noise. Additionally, SAM utilizes the age structure of the observed catch even in years when the overall catch value is considered biased. SAM was considered by recent benchmarks of North Sea cod (ICESWKCOD 2011; ICES-WKNSEA 2015) to be the most appropriate modelling approach for the stock assessment.

The assessement uses a time-varying maturity ogive, obtained by smoothing through an area-weighted maturity age key derived from the NS-IBTS Q1 survey data. The WG rejected the 2016 maturity estimates on the basis of low sample size and instead smoothed maturities to 2015 and assumed the 2015 maturity values to estimate SSB in 2016. This affects estimates of SSB, but has no impact in the assessment because recruitment is modelled as a random-walk process independent of SSB. An additional run of SAM using a maturity ogive smoothed to the 2016 maturity values is presented as a sensitivity check on SSB.

Figure 14.9a shows the SAM assessment with maturities smoothed to 2015, and Figure 14.9 b the SSB plot with maturities smoothed to 2016; the final assessment from last year (2015) is given in light grey for comparison. Rejecting the 2016 maturity estimates results in a SSB almost 15 thousand tonnes higher than had they been included, but has no further impact on the assessment.

Normalised residual plots are shown in Figure 14.10 , indicating no serious model misspecification. Retrospective plots for SSB, average fishing mortality and recruitment at age 1 are shown with Mohn's r statistics in Figure 14.11, indicating no serious retrospective patterns. A summary of the SAM final assessment run in terms of population
trends is provided in Figure 14.12, and the mean fishing mortality split into landings and discards, using landings fraction, and split into ages is shown in Figure 14.13.

### 14.3.4 Final assessment

The SAM update run with maturity smoothed to 2015 is accepted as the final assessment. The data used in the assessment are given in Tables 14.2-3 and 14.5-6, and the model configuration in Table 14.7a. Model fitting diagnostics, parameter estimates and associated correlation matrix are given in Table 14.7b, while normalised residual plots and retrospective runs are shown in Figures 14.10 and 14.11 respectively. Estimates of fishing mortality at age, stock numbers at age and total removals at age are given in Tables 14.8-10 respectively, while a summary table for estimates of recruitment (age 1), TSB, SSB, total removals and Fbar (2-4) are given in Table 14.11a (along with $95 \%$ confidence bounds), and estimates of landings, discards, catch, the catch multiplier and total removals (combining all these components) are given in Table 14.11b (and can be compared to the corresponding data in Table 14.4). Table 14.11c provides estimates of the catch multiplier along with $95 \%$ confidence bounds. Summary plots of the final assessment in terms of population trends is provided in Figure 14.12, and the mean fishing mortality split into landings and discards, using landings fraction, and split into age is shown in Figure 14.13. A comparison with last year's assessment is provided in Figure 14.14.

### 14.4 Historic Stock Trends

The historic stock and fishery trends are presented in Figures 14.12-13 and Table 14.11a-c.

Recruitment has fluctuated at a relatively low level since 1998. The 1996 year class was the last large year class that contributed to the fishery, and subsequent year classes have been the lowest in the time series apart from the 1999, 2005, 2009 and 2013 year classes. The 2006-8, 2010-12 and 2014 year classes are estimated to be weak.

Fishing mortality increased until the early 1980's, remained high until 2000 after which it has declined, and is now just below 0.4 , the target for the management plan when SSB $>$ Bap $_{\text {pa }}$.

SSB declined steadily during the 1970's and 80's. There was a small increase in SSB following improved recruitment coupled with a slight dip in fishing mortality in the mid 1990s, but with low recruitment since 1998 and continued high mortality rates, SSB continued to decline. SSB is estimated to have increased in recent years from the lowest level in the time series in 2006. TSB estimates have been increasing for slightly longer than SSB because of the 2005 year class, but have not experienced as rapid an improvement as SSB because of continued low recruitment.

Figure 14.15 indicates that the age structure in the population is gradually improving (number of fish aged 5 and older in the population appears to be increasing), and the survival of fish to age 5 is at its highest level in the time series.

Biomass indices by subregion (Figure 14.16a with subregions given in Figure 14.16c) highlight differing rates of change in cod biomass, with a general decline in all areas prior to the mid-2000s, and a general increase in all areas thereafter, apart from the southern area, where biomass has not increased following the decline. Recruitment indices by subregion (Figure 14.16b with subregions given in Figure 14.16c) show similar trends in all areas. Management measures ensuring sustainable exploitation of substocks may be needed in addition to management for the stock as a whole.

### 14.5 Recruitment estimates

Estimates of recruitment were sampled from the 1997-2014 year classes, reflecting recent low levels of recruitment, but including the stronger 1999, 2005, 2009 and 2013 year classes. These re-sampled recruitments are only used for SAM forecasts in order to evaluate future stock dynamics.

### 14.6 MSY estimation

MSY estimation is performed with the EQSIM software (ICES-WGMG 2013), in accordance with the guidelines provided in ICES-WKMSYREF3 (2014). MSY estimation for North Sea cod was performed during the WKMSYREF3 meeting in late 2014 (ICESWKMSYREF3 2014) and repeated during WKNSEA (ICES-WKNSEA 2015) and WGNSSK (ICES-WGNSSK 2015) in 2015; the Blim used in the analysis was taken as the SSB associated with the 1996 year class (the last reasonably-sized recruitment; Section 14.9). MSY ranges for NS cod were pulished during 2015 following an EU request to provide plausible values around FMSY (ICES-Special Request Advice 2015).

Assessment error in the advisory year and associated autocorrelation was derived from MSE evaluations of the current EU management plan for both assessments. There were three choices for recruitment periods, namely the full time series, only recruitment from 1988 onwards (reflecting the period of known productivity change in the North Sea), and only recruitment from 1998 onwards (reflecting the recent low period of recruitment for North Sea cod). The second of these (1988 onwards) was selected for the analysis because it was a period that included the SSB used for Blim, reflected the productivity change in the North Sea, and excluded the "gadoid outburst" of the 1960s and 1970s that could be considered an exception. Nevertheless, there are indications that recruitment from 1998 onwards has been lower than would be explained by SSB alone, so an EQSIM analysis based on the very low recruitment period of 1998 onwards was used as a precautionary check on the Fmsy range. Further investigation is needed to evaluate whether this very low recruitment period is just due to short-term environmental effects, or whether it is likely to continue in the long term; such changes may influence both the recovery rate of SSB and the values for biomass reference points.

A summary of the resultant biological reference points based on the recruitment period 1988-2014 (not including the advisory HCR in all but FP.05) is provided in the following table.

| Stоск |  |
| :--- | :--- |
| FMSY | 0.33 |
| FMSY lower | 0.22 |
| FMSY upper | 0.49 |
| FP. 05 (5\% risk to Blim, no HCR included) | 0.62 |
| FP.05 (5\% risk to Blim, with HCR included) | 0.75 |
| FMSY upper precautionary | $0.49^{*}$ |
| MSY | 102903 t |
| Median SSB at FMSY | 466778 t |
| Median SSB at FMSY upper precautionary | 351435 t |
| Median SSB at FMSY lower | 687971 t |

*Note that for the recruitment period 1998-2014, the Fp. 05 value is 0.42 for an EQSIM run with no HCR included, and 0.52 for an EQSIM run with HCR included, so in the case where the HCR is included, the Fmsy upper value is not constrained.

### 14.7 Short-term forecasts

## The May forecast

Forecasting takes the form of short-term stochastic projections. A total of 1000 samples are generated from the estimated distribution of survivors, with recruitment being sampled with replacement from the year 1998 to the final year of catch data (a period during which recruitment has been low). These replicates are then simulated forward according to model and forecast assumptions (see Table below), using the usual exponential decay equations, but also incorporating the stochastic survival process (using the estimated survival standard deviation) and subject to different catch-options scenarios.

Forecasts are presented for the final SAM run with maturities smoothed to 2015 (Table 14.12a) and for the sensitivity run with maturites smoothed to 2016 (Table 14.12b). Forecast assumptions are as follows. [Note that the values that appear in the catch options Tables 14.12a and b are medians from the distributions that result from the stochastic forecast.]

| Initial stock size | Starting populations are simulated from the estimated distribution <br> at the start of the intermediate year (including co-variances). |
| :--- | :--- |
| Maturity | Maturity for the intermediate year is taken from the smoothed <br> maturity ogive. Maturity for the TAC year onwards is the average of <br> final three years of assessment data |
| Natural mortality | Average of final three years of assessment data. |
| F and M before spawning | Both taken as zero. |
| Weight at age in the catch | Average of final three years of assessment data. |
| Weight at age in the stock | Assumed to be the same as weight at age in the catch. |
| Exploitation pattern | Fishing mortalities taken as a three year average scaled to the final <br> year. |
| Intermediate year <br> assumptions | Multiplier reflecting intended changes in effort (and therefore F) <br> relative to the final year of the assessment, assumed to be 1 to reflect <br> a status quo intermediate year assumption. |
| Stock recruitment model <br> used | Recruitment for the intermediate year onwards (the year the WG <br> meets) is sampled, with replacement, from 1998 to the final year of <br> catch data. |
| Procedures used for | The final year landing fractions are used in the forecast period. <br> splitting projected catches |

Large differences in SSB in the intermediate year between the assessment and the forecast resulted from assuming the final year of assessment data in 2016 for the assessment and the average of the final three years assessment data in 2016 for the forecast. This divergence had not presented itself to this extent in previous years, and was solved by using three year averages for stock weights and natural mortality to calculate SSB in the intermediate year for the assessment, and by using the smoothed maturity estimates to calculate SSB in the intermediate year for the forecast (which were set equal to 2015 maturity estimates in the final assessment) so the two calculations are consistent.

This is the first year that maturity data has been available in the assessment year and, for the sensitivity run where the 2016 maturity estimates were retained, necessitated increasing the forecast assumption for maturity from a three to a four year average. This is consistent with the start of the period over which the other data are averaged and allows inclusion of the most recent maturity estimate.

## The October forecast

Since the NS-IBTS Q3 index has been re-introduced into the assessment, there is an opportunity to update the forecast in October following the NS-IBTS Q3 survey. ICESWKNSEA (2015) recommended that the usual procedure be used to establish whether to re-open advice in the autumn (as described in ICES-AGCREFA 2008). Once it has been established that advice should be re-opened for North Sea cod, the recommended procedure is to then re-run the assessment and forecast with the new Q3 data included, but to use the actual SAM estimate of recruitment for the intermediate year (the year following the final year of catch data), with recruitment for the years following the intermediate year being re-sampled, with replacement, from the period 1998 to the final year of catch data.

The ICES-WKNSEA (2015) recommendations on conducting the North Sea cod forecast deviated from the ICES norm in that the October forecast implies re-running the SAM assessment, and was therefore presented to the ICES ACOM leadership who have given it their approval. The forecasting procedure will therefore follow the ICESWKNSEA (2015) recommended approach.

## The current May forecast

A number of scenarios were considered as follows [note, $\mathrm{B}_{\text {trigger }}=\mathrm{B}_{\mathrm{pa}}=165000 \mathrm{t}$, and $\mathrm{F}_{\mathrm{ms}}=0.33$; see Section 14.9]:

1. EU Management plan: the Longterm Phase of the plan, applying the sliding rule with former $B_{\lim }$ and $B_{p a}$ values ( $70000 t$ and $150000 t$ respectively) (paragraph 4 of Article 8 of EC 1342/2008), ensuring that TAC (2017) is within 20\% of TAC (2016)
2. EU-Norway agreement plan: the Longterm Phase of the plan, applying the same sliding rule as for the EU Management plan, but using the new $\mathrm{Blim}_{\text {lim }}$ and $B_{\text {pa }}$ values ( 118000 t and 165000 t respectively) (see Section 14.9), ensuring that TAC (2017) is within $20 \%$ of TAC (2016)
3. $\operatorname{MSY}$ framework: $\operatorname{Fbar}(2017)=F_{M S Y} \times \min \left\{1 ; S S B_{2017} / B_{t r i g g e r}\right\}$
4. Zero catch: Fbar $(2017)=0$
5. MSY: Fbar (2017) = FMSY
6. $\quad \mathrm{F}_{\mathrm{pa}}: \operatorname{Fbar}(2017)=\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} / 1.4=0.41$
7. $\mathrm{F}_{\mathrm{pa}}: \operatorname{Fbar}(2017)=\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\lim } \times \exp (-\sigma \times 1.645)=0.47$
8. Flim: Fbar $(2017)=$ Flim
9. $\operatorname{SSB}(2018)=$ Blim: F corresponding to SSB $(2018)=B_{\lim }$
10. $\operatorname{SSB}(2018)=B_{\mathrm{pa}}:$ F corresponding to $\operatorname{SSB}(2018)=\mathrm{B}_{\mathrm{pa}}$
11. SSB (2018) $=B_{\text {trigger: }}$ F corresponding to SSB $(2018)=B_{\text {trigger }}$
12. Lower TAC constraint: Fbar (2017) such that TAC (2017) $=0.8 \times$ TAC (2016)
13. Rollover TAC - 15\%: Fbar (2017) such that TAC (2016) $=0.85 \times$ TAC (2016)
14. Rollover TAC - 10\%: Fbar (2017) such that TAC (2017) $=0.9 \times$ TAC (2016)
15. Rollover TAC - 5\%: Fbar (2017) such that TAC (2017) $=0.95 \times$ TAC (2016)
16. Rollover TAC: Fbar (2017) such that TAC (2017) = TAC (2016)
17. Rollover TAC $+5 \%$ : Fbar (2017) such that TAC $(2017)=1.05 \times$ TAC (2016)
18. Rollover TAC $+10 \%$ : Fbar (2017) such that TAC $(2017)=1.1 \times$ TAC (2016)
19. Rollover TAC $+15 \%$ : Fbar (2017) such that TAC $(2017)=1.15 \times$ TAC (2016)
20. Upper TAC constraint: Fbar (2017) such that TAC (2017) $=1.2 \times$ TAC (2016)
21. Status quo - constant F: Fbar (2016) $=$ Fbar (2015)

The reason two management plan options ( 1 and 2 above) are supplied is because both plans were based on $B_{l i m}$ and $B_{p a}$ as part of the sliding rule, but with the revision of these reference points in 2015, the two plans now differ from one another. The EU management plan continues to be based on the previous values for $B_{\lim }$ and $B_{p a}$ (formerly 70000 t and 150000 t respectively) while the EU-Norway agreement has the flexibility to accommodate the revised values for these quantities (118000t and 165000 t respectively). Furthermore, both management plans switched into their long-term phases (when they were still based on the same values for $B_{\lim }$ and $B_{p a}$ ) in 2013.

Two catch options ( 6 and 7 above) are supplied for $\mathrm{F}_{\mathrm{pa}}$. The first relates to the value of $\mathrm{F}_{\mathrm{pa}}$ derived from $\mathrm{F}_{\text {lim }} / 1.4$ while the second relates to the $\mathrm{F}_{\mathrm{pa}}$ value based on assessment uncertainty in fishing mortality in the terminal year (Section 14.9).

Forecasts for the SAM final run and associated scenarios are given in Table 14.12a. For completeness, Table 14.12b provides the corresponding forecasts for the SAM sensitivity run with maturities smoothed to 2016, excluding options 8-11.

### 14.8 Medium-term forecasts

Medium-term projections are not carried out for this stock.

### 14.9 Biological reference points

Biological reference points were calculated in 2015 on the basis of the SAM final assessment (ICES-WGNSSK 2015). The choice for Blim was to take the last SSB to have produced a reasonably-sized recruitment (the 1996 year class); the reason the changepoint in a segemented regression fitted to the whole time period was not used was because this time period spans different environmental and recruitment regimes, and such a changepoint would therefore not be appropriate for deriving $\mathrm{Blim}_{\mathrm{lim}}$. The SSB in 1996 produced the last outstanding year class (1162 403 thousands recruits based on the 2015 assessment) that was above the average observed between 1963 and 1996 (1 029484 thousands based on the 2015 assessment) when the stock produced relatively high recruitment compared to recently observed values. Therefore, it can be argued that a SSB above the one observed in 1996 has the potential to produce high recruitment under sufficiently good environmental conditions, and therefore impaired recruitment because of a too-low SSB is avoided. Bpa was simply calculated as $1.4 \times$ Blim.

Flim was calculated on the basis of the SAM final assessment from 2015, for consistency with existing reference points. Flim estimation was performed with the EQSIM software on the basis of the very low recruitment period from 1998, consistent with the calcualtion of $\mathrm{Fp}_{\mathrm{p} .05}$ used as a precautionary check on the $\mathrm{F}_{\text {msy }}$ range (as opposed to the period 1988 onwards used for calculation of $\mathrm{Blim}_{\mathrm{lim}}$ and $\mathrm{B}_{\mathrm{pa}}$ ). The changepoint of the segmented regression was estimated rather than forced at $\mathrm{B}_{\text {lim. This deviation from the ICES guide- }}$ lines avoids use of a stock-recruit curve that falls below the majority of observed stockrecruit pairs and is consistent with the curve used for calculating the existing Fr. 05 value. $\mathrm{F}_{\mathrm{pa}}$ was simply calculated as $\mathrm{F}_{\mathrm{lim}} / 1.4$, and an alternative $\mathrm{F}_{\mathrm{pa}}$ value was calculated on the
basis of assessment uncertainty in fishing mortality in the final assessment year ( $\sigma=$ $0.12 ; \mathrm{F}_{\mathrm{pa}}=0.47$ ) following the ICES guidelines. Biological reference points are as follows:

| Framework | Reference POINT | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY <br> Btrigger | 165000 t . | The default option of Bpa.(=1.4×Blim) |  |
|  | FMSY | 0.33 | EQSim analysis based on recruitment period 1988-2014 | 2015 assessment |
| Precautionary approach | Blim | 118000 t . | SSB associated with the 1996 year class | 2015 assessment |
|  | Bpa | 165000 t . | Blim multiplied by 1.4. This is the current ICES default approach. |  |
|  | Flim | 0.58 | EQSim analysis based on recruitment period 1998-2014 | 2015 assessment |
|  | Fpa | 0.41 | Flim/1.4 |  |
| EU <br> Management plan | SSBlower | 70000 t . | Former Blim |  |
|  | SSBupper | 150000 t | Former Bpa |  |
|  | Flower | 0.2 | Fishing mortality when SSB <SSBlower. | EC 1342/2008 |
|  | Fupper | 0.4 | Fishing mortality when SSB>SSBupper |  |
| EU-Norway agreement | SSBlower | 118000 t . | Revised Blim | 2008 EU- <br> Norway agreement |
|  | SSBupper | 165000 t | Revised Bpa |  |
|  | Flower | 0.2 | Fishing mortality when SSB <SSBlower. |  |
|  | Fupper | 0.4 | Fishing mortality when SSB>SSBupper |  |

### 14.10Quality of the assessment

The quality of the commercial landings and catch-at-age data for this stock deteriorated in the 1990s following reductions in the TAC without associated control of fishing effort. The WG considers the international landings figures from 1993 onwards to have inaccuracies that lead to retrospective underestimation of fishing mortality and over estimation of spawning stock biomass and other problems with an analytical assessment. The mismatch between reported and actual landings is assumed to be negligible since 2006.

Prior to 2002 estimates of discards for areas 4 and 7.d are taken from the Scottish discard sampling program and the average proportions across gears applied to raise the landings data from other areas. If the gear and fishery characteristics differ, this could introduce bias. This bias is likely to introduce sensitivity to the estimates of the youngest age classes (1 and 2) and will not affect estimates of SSB. Intercatch has been used to raise data for discards ratios and landings and discard age compositions from 2002 onwards. The provision of discard information has vastly improved since 2009.

Comparing the assessment this year with last year gives the following (Figure 14.14): historical SSB trends are similar; the stock is above $\mathrm{B}_{\mathrm{lim}}$ and approaching $\mathrm{B}_{\mathrm{pa}}$; fishing mortality continues to decline, and is now just below the management plan target of 0.4 , but still above Fmsy; there is hardly any difference in fishing mortality compared to last year.

The estimated CVs for observed catch at age 1, for the NS-IBTS Q1 and Q3 survey indices at age 1 and for the stock-recruitment relationship are all large: $61 \%, 46 \%, 36 \%$ and $75 \%$, respectively. These large CVs suggest that these sources of information are somewhat ignored in the SAM recruitment estimation, which might therefore be more influenced by age 2 abundance estimates and model assumptions about F-at-age 1. The CV of the survival process is assumed to be the same for all non-recruiting ages (estimated at $11 \%$ ) and this might have an impact on recruitment estimates (and, hence, age 1 catch and survey residuals) because it constraints the changes permitted between abundance at ages 1 and 2 of a cohort.

Finally, the high correlation (0.89) estimated for the increments of $\log [\mathrm{F}(\mathrm{y}, \mathrm{a})]$ across ages suggests that the model might react a bit slowly if different changes in selectivity start to happen for different ages (for example, as a consequence of discard reduction policies). Annual assessment results should be monitored closely, via retrospective analyses and other model diagnostics.

Changes to the assessment in 2015 include a reduction of the plusgroup from $7+$ to $6+$. This reduces the cohort information for ages $6+$; these ages represent around $26 \%$ of the SSB (by weight), and if the SSB continues to increase, this proportion should also increase as more fish aggregate in the plusgroup, with an associated increasing loss in cohort signal for ages in the plusgroup, potentially undermining the assessment. Furthermore, this change introduced increasingly domed selection in the latter half of the time series that was not present in previous assessments; although there are reasons why such increasingly domed selection might occur, such as some evidence that larger cod inhabit less accessible rocky areas or simply move away from areas fishing vessels operate in, these reasons remain largely speculative.

The SAM model estimates the quantity of additional "unaccounted removals" that would be required to be added or removed from the catch-at-age data in order to remove any persistent trends in survey catchability. The unaccounted removals figures given by SAM could potentially include components due to increased natural mortality and discarding as well as misreported landings.
There is general agreement across all models presented (SAM and SURBAR) of an increasing SSB since the mid-2000s, declining fishing mortality (total mortality for SURBAR) since around 2000, and slightly stronger 2005, 2009 and 2013 year classes in recent years. The decline in fishing mortality is evident from the shallower gradients of logcatch curves, and the stronger 2013 year class is evident from this year class being slightly more widespread in the North Sea compared to other recent year classes at the same age.

The annually varying maturity-at-age estimates are derived from an area-weighted maturity age key based on NS-IBTS Q1 data from the period 1977-2015, to which a smoother is applied to get rid of the effects of variations in sampling intensity. Maturity sampling in the southern North Sea was poor in 2016, leading the WG to reject the 2016 maturity estimates. This increased the estimate of SSB in 2016 by almost 15 thousand tonnes but had no further impact on the assessment.

Values for natural mortality were updated in 2015, following the key run conducted by WGSAM (ICES-WGSAM, 2014); they are smoothed annual model estimates from a multi-species model. A Delta-GAM approach, assuming a stationary spatial model with ship effect, has been used to derive both Q1 and Q3 NS-IBTS indices.

### 14.11 Status of the Stock

There has been a strong improvement of the status of the stock in the last few years. SSB has increased from the historical low in 2006, and is now above Blim and approaching $B_{p a}$. This increasing trend is expected to continue in the short term under current fishing mortality levels, because of improved survival of incoming year classes.

Fishing mortality has declined from 2000, and is now just below the target for the management plan when $S S B>B_{p a}$, but still estimated to be above the level that achieves the long-term objective of maximum yield.

Recruitment of 1 year old cod has varied considerably since the 1960s, but since 1998, average recruitment has been lower than any other time. The 2009 and 2013 year classes are stronger but the 2010-12 and 2014 year classes appear to be weak, with some indication (based on one survey only) of a weak 2015 year class. Recent sharp increases in the rate of discarding have been reversed are are stabilising at lower levels.

### 14.12 Management Considerations

The stock has begun to recover from the low levels to which it was reduced in early 2000, at which recruitment was impaired and the biological dynamics of the stock difficult to predict. Fishing mortality rates have been reduced from 2000 and in combination with the stronger 2005, 2009 and 2013 year classes, the stock has increased since 2006. The reduction in fishing mortality is allowing the recent series of poor recruitments to make an improved contribution to the stock. The low average age of the spawning stock reduces its reproductive capacity as first-time spawners reproduce less successfully than older fish, a factor that has contributed to the continued low recruitment.

There may have been some difficulties with the effectiveness of the cod recovery plans; despite the objective to reduce fishing mortality and to increase the SSB by combined TAC control and effort management, estimated total removals have until recently been much higher than intended. The situation has been improving, however, and fishing mortality is now just below the management plan target when $\mathrm{SSB}>\mathrm{B}_{\mathrm{pa}}$. Discarding currently contributes around a quarter of the total catch by weight, a substantial improvement compared to recent years (when the average was almost half of the total). There have been considerable efforts to reduce discards by some countries, and the impact of these reductions are starting to be felt (e.g. reduced discarding leading to improved survival of incoming year classes).

Rejecting the 2016 maturity estimates has no impact on the catch advice when following the ICES MSY approach as SSB in 2017 is above MSY B trigger for both assessment runs. It does however impact the management plans because rejecting the 2016 maturity estimates results in an intermediate SSB that is either above or closer to $\mathrm{B}_{\mathrm{pa}}$ defined in the management plans, which then impacts the F set following the sliding rule.

Cod is caught by a large variety of gears and together with many other species. It is important to consider both the species-specific assessments of these species for effective management, but also the broader mixed-fisheries context. This is not straightforward when stocks are managed via a series of single-species management plans that do not incorporate such mixed-stocks considerations. However, a reduction in effort on one stock may lead to a reduction or an increase in effort on another, and the implications of any change need to be considered carefully. The ICES WGMIXFISH Group
monitors the consistency of the various single-species management plans under current effort schemes, in order to estimate the potential risks of quota over- and undershooting for the different stocks.

There is a need to reduce fishing induced mortality on North Sea cod further, particularly for younger ages, in order to allow more fish to reach maturity and increase the probability of good recruitment. Incidence of discarding remain high, with the proportion of fish discarded by number in 2015 being $86 \%$ of 1 year old (compared to $78 \%$ in 2014 ), $58 \%$ of 2 year old ( $61 \%$ in 2014), $29 \%$ of 3 year old ( $24 \%$ in 2014) and $8 \%$ of 4 year old cod (9\% in 2014).

Because the fishery is at present so dependent on incoming year classes, fishing mortalities on these year classes remain high, and only a small proportion of 2 year olds currently survive to maturity. At the same time, the unbalanced age structure of the stock reduces its reproductive capacity even if a sufficient SSB were reached, as firsttime spawners reproduce less successfully than older fish. Both factors are believed to have contributed to the reduction in recruitment of cod. However, there are indications that, although still low, survival to age 5 is improving, and is currently at the highest level in the time series.

The recruitment of the relatively more abundant year classes to the fishery may have no beneficial effect on the stock if they are caught and heavily discarded. The last substantial year class to enter the fishery was the 1996 year class. This year class was a prominent feature in all surveys, was heavily exploited and discarded by the fishery at ages $1-5$, and disappeared relatively quickly from the fishery. Furthermore, the 2014 years class is estimated to be weak, and there are some indications, based on one survey only, of a weak 2015 year class.

The availability of discard rate estimates has vastly improved since 2009, and catch estimates (landings and discards) are now provided by Intercatch from 2002 onwards.

Recent measures to improve survival of young cod, such as the Scottish Credit Conservation Scheme, and increased uptake of more selective gear such as the now widespread use of sorting grids in the Skagerrak, should be encouraged.

The reported landings in 2015 were 37.2 thousand tonnes and the estimated discards in 2015 were 12.6 thousand tonnes, giving a total of 49.8 thousand tonnes. Cod are taken by towed gears in mixed demersal fisheries, which include haddock, whiting, Nephrops, plaice, and sole. They are also taken in directed fisheries using fixed gears.

Cod catch in Division d was previously managed by a TAC for Divisions b-k, 8, 9,10, and CECAF 34.1.1, (i.e. the TAC covered a small proportion of the North Sea cod stock together with cod in Divisions7.e-k). Division 7.d is now allocated a separate TAC (since 2009), which is adjusted in line with the revision to the North Sea TAC.

### 14.13Assessment frequency

The frequency of assessments was discussed at the ACOM December 2014 meeting and the Committee decided to develop simple criteria to be used to identify stocks that would be candidates for less frequent assessments. A set of four criteria were suggested based on (1) the life span of the stock, (2) stock status, (3) relative importance of recruitment in the catch forecast and (4) the quality of the assessment. The assessment failed to meet two of the four criteria (highlighted in grey). Therefore the North Sea cod stock is not a candidate for less frequent assessments.

| Criterion |  |
| :---: | :---: |
| (1) Life span (i.e. maximum normal age) of the species is larger than 5 years | Life span larger than 5 years |
| (2) The stock status in relation to the reference points is according to the MSY criteria F (latest assessment year) is <= Fupper (upper bound in F range) AND SSB (start of intermediate year) >= MSY Btrigger | $\begin{aligned} & \mathrm{F}(2015)=0.385<\text { Fupper }=0.42 \\ & \mathrm{SSB}(2016)=161135<\text { Btrigger }=165000 \end{aligned}$ |
| (3) The average contribution to the catch in numbers of the recruiting year class in latest 5 years is less than $25 \%$ of the total catch in numbers. | The average contribution to the catch in numbers of the recruiting year class in latest 5 years is $28 \%$ of the total catch in numbers |
| (4) The retrospective pattern, based on a seven years peel of Mohn's rho index, shows that F is consistently underestimated by less than $20 \%$ | $\varrho=-0.035$ <br> i.e. F is overestimated by $4 \%$ |

### 14.14References

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Table 14.1 Nominal landings (in tons) of COD in 3.a (Skagerrak), 4 and 7.d, as officially reported to ICES, and as used by the Working Group.

Table 14.1.

| Sub-area IV |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| Belgium | 3,458 | 4,642 | 5,799 | 3,882 | 3,304 | 2,470 | 2,616 | 1,482 | 1,627 | 1,722 |
| Denmark | 23,573 | 21,870 | 23,002 | 19,697 | 14,000 | 8,358 | 9,022 | 4,676 | 5,889 | 6,291 |
| Faroe Islands | 44 | 40 | 102 | 96 | - | 9 | 34 | 36 | 37 | 34 |
| France | 1,934 | 3,451 | 2,934 | . | 1,222 | 717 | 1,777 | 620 | 294 | 664 |
| Germany | 8,344 | 5,179 | 8,045 | 3,386 | 1,740 | 1,810 | 2,018 | 2,048 | 2,213 | 2,648 |
| Greenland |  |  |  |  |  |  |  |  |  | 35 |
| Netherlands | 9,271 | 11,807 | 14,676 | 9,068 | 5,995 | 3,574 | 4,707 | 2,305 | 1,726 | 1,660 |
| Norway | 5,869 | 5,814 | 5,823 | 7,432 | 6,410 | 4,369 | 5,217 | 4,417 | 3,223 | 2,900 |
| Poland | 18 | 31 | 25 | 19 | 18 | 18 | 39 | 35 | - | - |
| Sweden | 617 | 832 | 540 | 625 | 640 | 661 | 463 | 252 | 240 | 319 |
| UK (E/W/NI) | 15,930 | 13,413 | 17,745 | 10,344 | 6,543 | 4,087 | 3,112 | 2,213 | 1,890 | 1,270 |
| UK (Scotland) | 35,349 | 32,344 | 35,633 | 23,017 | 21,009 | 15,640 | 15,416 | 7,852 | 6,650 | 4,936 |
| UK (combined) | n/a | $\mathrm{n} / \mathrm{a}$ | n/a | n/a | n/a | n/a | n/a | n/a | n/a | n/a |
| Others | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Danish industrial by-catch * |  | . | . | . | . | . | 105 | 22 | 17 | 21 |
| Norwegian industrial by-catch |  |  |  |  |  |  |  |  |  |  |
| Total Nominal Catch | 104,407 | 99,423 | 114,324 | 77,566 | 60,881 | 41,713 | 44,526 | 25,958 | 23,806 | 22,500 |
| Unallocated landings | 2,161 | 2,746 | 7,779 | 826 | -1,114 | -740 | -226 | -111 | -1,277 | 356 |
| WG estimate of total landings | 106,568 | 102,169 | 122,103 | 78,392 | 59,767 | 40,973 | 44,300 | 25,847 | 22,529 | 22,855 |
| Agreed TAC | 130,000 | 115,000 | 140,000 | 132,400 | 81,000 | 48,600 | 49,300 | 27,300 | 27,300 | 27,300 |
| Division VIld |  |  |  |  |  |  |  |  |  |  |
| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| Belgium | 321 | 310 | 239 | 172 | 110 | 93 | 51 | 54 | 47 | 51 |
| Denmark | - | - | - | - | - | - | - | - | - | - |
| France | 2,808 | 6,387 | 7,788 |  | 3,084 | 1,677 | 1,361 | 1,730 | 810 | 986 |
| Netherlands | - | - | 19 | 3 | 4 | 17 | 6 | 36 | 14 | 9 |
| UK (E/W/NI) | 414 | 478 | 618 | 454 | 385 | 249 | 145 | 121 | 103 | 184 |
| UK (Scotland) | 4 | 3 | 1 | - | - | - | - | - | - | - |
| UK (combined) | n/a | n/a | n/a | n/a | n/a | n/a | n/a | n/a | n/a | n/a |
| Total Nominal Catch | 3,547 | 7,178 | 8,665 | 629 | 3,583 | 2,036 | 1,563 | 1,941 | 974 | 1,230 |
| Unallocated landings | -44 | -135 | -85 | 6,229 | -1,258 | -463 | 1,534 | -707 | 40 | 29 |
| WG estimate of total landings | 3,503 | 7,043 | 8,580 | 6,858 | 2,325 | 1,573 | 3,097 | 1,234 | 1,014 | 1,259 |
| Division Illa (Skagerrak)** |  |  |  |  |  |  |  |  |  |  |
| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| Denmark | 14,573 | 12,159 | 12,339 | 8,681 | 7,684 | 5,900 | 5,525 | 3,067 | 3,038 | 3,019 |
| Germany | 259 | 81 | 54 | 54 | 54 | 32 | 83 | 49 | 99 | 86 |
| Norway | 1,046 | 1,323 | 1,293 | 1,146 | 926 | 762 | 645 | 825 | 856 | 759 |
| Sweden | 1,986 | 2,173 | 1,900 | 1,909 | 1,293 | 1,035 | 897 | 510 | 495 | 488 |
| Others | - | - | - | - | - | - | - | 27 | 24 | 21 |
| Danish industrial by-catch * | 676 | 205 | 97 | 62 | 99 | 687 | 20 | 5 | 4 | 2 |
| Total Nominal Catch | 17,864 | 15,736 | 15,586 | 11,790 | 9,957 | 7,729 | 7,170 | 4,483 | 4,516 | 4,375 |
| Unallocated landings | -1,615 | -790 | -255 | -816 | -680 | -643 | 298 | -692 | -602 | -376 |
| WG estimate of total landings | 16,249 | 14,946 | 15,331 | 10,974 | 9,277 | 7,086 | 7,468 | 3,791 | 3,914 | 3,998 |
| Agreed TAC | 23,000 | 16,100 | 20,000 | 19,000 | 11,600 | 7,000 | 7,100 | 3,900 | 3,900 | 3,900 |
| Sub-area IV, Divisions VIld and Illa (Skagerrak) combined |  |  |  |  |  |  |  |  |  |  |
|  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| Total Nominal Catch | 125,818 | 122,337 | 138,575 | 89,985 | 74,421 | 51,478 | 53,260 | 32,382 | 29,296 | 28,104 |
| Unallocated landings | 502 | 1,821 | 7,439 | 6,240 | -3,052 | -1,846 | 1,605 | -1,510 | -1,839 | 9 |
| WG estimate of total landings | 126,320 | 124,158 | 146,014 | 96,225 | 71,369 | 49,632 | 54,865 | 30,872 | 27,457 | 28,113 |
| ** Skaggerak/Kattegat split derived from national statistics |  |  |  |  |  |  |  |  |  |  |
| * The Danish industrial by-catch (up to 2001) are not included in the (WG estimate of) total landings |  |  |  |  |  |  |  |  |  |  |
| . Magnitude not available - Magnitude known to be nil <0.5 Magnitude less than half the unit used in the table n /a Not applicable |  |  |  |  | half the unit used in the table n/a Not applicable |  |  |  |  |  |
| Division IV and Illa (Skagerrak) landings not included in the assessment |  |  |  |  |  |  |  |  |  |  |
| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| Danish industrial by-catch * | 676 | 205 | 97 | 62 | 99 | 687 | - | - | - | - |
| Norwegian industrial by-catch |  |  |  |  |  |  |  | - |  |  |
| Total | 676 | 205 | 97 | 62 | 99 | 687 | 0 | 0 | 0 | 0 |

Table 14.1 cont. Nominal landings (in tons) of COD in 3.a (Skagerrak), 4 and 7.d, as officially reported to ICES, and as used by the Working Group.

Table 14.1. Cont'd.

| Sub-area IV |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Country | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| Belgium | 1,309 | 1,009 | 894 | 946 | 666 | 653 | 862 | 1,076 | 1,257 | 1,187 |
| Denmark | 5,105 | 3,430 | 3,831 | 4,402 | 5,686 | 4,863 | 4,803 | 4,536 | 5,457 | 6,026 |
| Faroe Islands | 3 | 0 | 16 | 45 | 32 | 0 | 0 | 0 | 0 |  |
| France | 354 | 659 | 573 | 950 | 781 | 619 | 368 | 287 | 638 | 521 |
| Germany | 2,537 | 1,899 | 1,736 | 2,374 | 2,844 | 2,211 | 2,385 | 1,921 | 2,257 | 2,133 |
| Greenland | 23 | 17 | 17 | 11 | 0 | 0 | 0 | 0 | 0 |  |
| Netherlands | 1,585 | 1,523 | 1,896 | 2,649 | 2,657 | 1,928 | 1,955 | 1,344 | 1,242 | 1,349 |
| Norway | 2,749 | 3,057 | 4,128 | 4,234 | 4,496 | 4,898 | 4,601 | 4,079 | 4,590 | 5,486 |
| Poland | 0 | 1 | 2 | 3 | 0 | 2 | 0 | 0 | 0 |  |
| Sweden | 309 | 387 | 439 | 378 | 363 | 315 | 472 | 332 | 401 | 417 |
| UK (E/W/NI) | 1,491 | 1,588 | 1,546 | 2,384 | 2,553 | 2,169 | 1,630 | 2,129 | 2,963 |  |
| UK (Scotland) | 6,857 | 6,511 | 7,185 | 9,052 | 11,567 | 10,141 | 10,565 | 10,619 | 10,517 |  |
| UK (combined) | n/a | n/a | $\mathrm{n} / \mathrm{a}$ | n/a | $\mathrm{n} / \mathrm{a}$ | n/a | n/a | n/a | 13,480 | 14,839 |
| Others | 786 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Danish industrial by-catch | 11 | 23 | 1 | 72 | 12 | 0 | 0 | 2 | 24 | 0 |
| Norwegian indust by-catch * | 48 | 101 | 22 | 4 | 201 | 1 |  |  | . |  |
| Total Nominal Catch | 23,119 | 20,104 | 22,264 | 27,500 | 31,657 | 27,799 | 27,641 | 26,325 | 29,346 | 31,959 |
| Unallocated landings | -2,041 | -1,047 | -607 | 134 | -677 | -1,124 | -1,014 | -1,010 | -796 | -715 |
| WG estimate of total landings | 21,078 | 19,056 | 21,657 | 27,634 | 30,980 | 26,675 | 26,627 | 25,315 | 28,550 | 31,244 |
| Agreed TAC | 23,205 | 19,957 | 22,152 | 28,798 | 33,552 | 26,842 | 26,475 | 26,475 | 27,799 | 29,189 |
| Division VIld |  |  |  |  |  |  |  |  |  |  |
| Country | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| Belgium | 80 | 84 | 154 | 73 | 57 | 56 | 40 | 53 | 72 | 79 |
| Denmark | - |  |  |  |  |  |  |  | . |  |
| France | 1,124 | 1,743 | 1,326 | 1,779 | 1,606 | 1,078 | 885 | 768 | 1,270 | 1,100 |
| Netherlands | 9 | 59 | 30 | 35 | 45 | 51 | 40 | 38 | 50 | 47 |
| UK (E/W/NI) | 267 | 174 | 144 | 133 | 127 | 125 | 99 | 100 | 156 |  |
| UK (Scotland) | 1 | 12 | 7 | 3 | 1 | 1 | 0 | 0 | 0 |  |
| UK (combined) | n/a | n/a | n/a | n/a | n/a | n/a | n/a | n/a | 156 | 161 |
| Total Nominal Catch | 1,481 | 2,072 | 1,661 | 2,023 | 1,836 | 1,311 | 1,064 | 959 | 1,548 | 1,387 |
| Unallocated landings | -2 | 75 | -32 | -136 | -128 | 8 | 56 | -43 | -112 | 11 |
| WG estimate of total landings | 1,479 | 2,147 | 1,629 | 1,887 | 1,708 | 1,319 | 1,120 | 916 | 1,436 | 1,398 |
| Agreed TAC |  |  |  | 1,678 | 1,955 | 1,564 | 1,543 | 1,543 | 1,620 | 1,701 |
| Division Illa (Skagerrak)** |  |  |  |  |  |  |  |  |  |  |
| Country | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| Denmark | 2,513 | 2,246 | 2,553 | 3,024 | 3,286 | 3,118 | 3,178 | 3,033 | 3,430 | 3,344 |
| Germany | 84 | 67 | 52 | 55 | 56 | 60 | 78 | 69 | 84 | 87 |
| Norway | 628 | 681 | 779 | 440 | 375 | 421 | 615 | 575 | 528 | 499 |
| Sweden | 372 | 370 | 365 | 459 | 458 | 518 | 520 | 529 | 570 | 576 |
| Others | 373 | 385 | 13 | 2 | 26 | 0 | 0 | 33 | 28 | 24 |
| Danish industrial by-catch | 3 | 2 | 7 | 2 | 10 | 0 | 1 | 1 | 5 | 5 |
| Total Nominal Catch | 3,973 | 3,751 | 3,769 | 3,982 | 4,211 | 4,117 | 4,392 | 4,240 | 4,645 | 4,536 |
| Unallocated landings | -715 | -731 | -376 | -188 | -154 | -161 | -65 | -86 | 42 | 27 |
| WG estimate of total landings | 3,258 | 3,020 | 3,393 | 3,794 | 4,057 | 3,956 | 4,327 | 4,154 | 4,687 | 4,563 |
| Agreed TAC | 3,315 | 2,851 | 3,165 | 4,114 | 4,793 | 3,835 | 3,783 | 3,783 | 3,972 | 4,171 |
| Sub-area IV, Divisions VIld and Illa (Skagerrak) combined |  |  |  |  |  |  |  |  |  |  |
|  | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| Total Nominal Catch | 28,573 | 25,927 | 27,694 | 33,505 | 37,705 | 33,227 | 33,097 | 31,524 | 35,538 | 37,882 |
| Unallocated landings | -2,759 | -1,704 | -1,015 | -190 | -959 | -1,277 | -1,023 | -1,139 | -865 | -676 |
| WG estimate of total landings | 25,815 | 24,223 | 26,679 | 33,315 | 36,746 | 31,950 | 32,074 | 30,386 | 34,673 | 37,205 |
| ** Skaggerak/Kattegat split derived from national statistics |  |  |  |  |  |  |  |  |  |  |
| * The Norwegian industrial by-catch are not included in the (WG estimate of) total landings |  |  |  |  |  |  |  |  |  |  |
| . Magnitude not available - Magn | - Magnitude known to be nil |  | <0.5 Magnitude less than half the unit used in the table $\mathrm{n} / \mathrm{a}$ Not applicable |  |  |  |  |  |  |  |
| Division IV and Illa (Skagerrak) landings not included in the assessment |  |  |  |  |  |  |  |  |  |  |
| Country | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| Danish indust by-catch | - | - | - | - | - | - | - | - | - | - |
| Norwegian indust by-catch * | 48 | 101 | 22 | 4 | 201 | 1 |  |  |  |  |
| Total | 48 | 101 | 22 | 4 | 201 | 1 | - | - | - | - |

Table 14.2a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Landings numbers at age (Thousands).

| Landings numbers at age (thousands) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE/YEAR | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 |
| 1 | 3198 | 5004 | 15734 | 18133 | 10749 | 5800 | 2932 | 54219 | 44599 | 3813 | 25836 |
| 2 | 42377 | 22373 | 51628 | 62202 | 70539 | 83416 | 22561 | 33747 | 154565 | 186744 | 31596 |
| 3 | 6995 | 20003 | 17557 | 29695 | 32529 | 42373 | 31419 | 18395 | 17132 | 47885 | 54655 |
| 4 | 3519 | 4285 | 9135 | 6153 | 11205 | 12330 | 13641 | 13272 | 6720 | 5653 | 14002 |
| 5 | 2774 | 1908 | 2375 | 3362 | 3255 | 6046 | 4542 | 6266 | 7065 | 2713 | 2195 |
| 6 | 1207 | 1809 | 946 | 1272 | 1964 | 1407 | 2881 | 1754 | 2686 | 3184 | 1103 |
| 7 | 81 | 596 | 655 | 475 | 884 | 866 | 585 | 956 | 888 | 1671 | 1055 |
| 8 | 489 | 117 | 297 | 368 | 353 | 307 | 420 | 208 | 455 | 609 | 487 |
| 9 | 13 | 93 | 51 | 125 | 137 | 150 | 147 | 185 | 227 | 388 | 79 |
| 10 | 6 | 11 | 75 | 56 | 40 | 111 | 46 | 97 | 77 | 112 | 57 |
| +gp | 0 | 4 | 8 | 83 | 17 | 24 | 77 | 40 | 93 | 17 | 161 |
| TOTALNUM | 60659 | 56203 | 98460 | 121923 | 131671 | 152829 | 79251 | 129139 | 234508 | 252789 | 131226 |
| TONSLAND | 115893 | 125393 | 180120 | 220197 | 251687 | 286948 | 199746 | 224993 | 326492 | 352161 | 237874 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| 1 | 15484 | 33210 | 5695 | 75130 | 29593 | 34627 | 62394 | 20131 | 66220 | 25488 | 64358 |
| 2 | 58624 | 46907 | 99779 | 50926 | 174912 | 91143 | 104356 | 187626 | 64755 | 128396 | 66026 |
| 3 | 11347 | 18849 | 18481 | 25525 | 17178 | 44384 | 34938 | 34567 | 59907 | 21456 | 31087 |
| 4 | 15745 | 4640 | 6707 | 4597 | 9396 | 4011 | 12274 | 8953 | 9487 | 11787 | 4238 |
| 5 | 4601 | 7525 | 1732 | 2286 | 2989 | 3375 | 1958 | 4088 | 3447 | 2803 | 3415 |
| 6 | 956 | 2057 | 3056 | 833 | 1103 | 708 | 1269 | 779 | 2048 | 1246 | 1013 |
| 7 | 436 | 447 | 920 | 1140 | 408 | 396 | 494 | 599 | 425 | 589 | 434 |
| 8 | 393 | 195 | 130 | 370 | 403 | 139 | 197 | 133 | 234 | 179 | 243 |
| 9 | 330 | 228 | 67 | 262 | 152 | 157 | 73 | 64 | 77 | 89 | 59 |
| 10 | 80 | 95 | 63 | 26 | 36 | 42 | 55 | 36 | 27 | 28 | 44 |
| +gp | 188 | 63 | 43 | 96 | 44 | 17 | 25 | 21 | 16 | 23 | 19 |
| TOTALNUM | 108183 | 114215 | 136672 | 161191 | 236214 | 178997 | 218034 | 256998 | 206643 | 192083 | 170937 |
| TONSLAND | 213215 | 204249 | 233007 | 208318 | 294640 | 266019 | 293753 | 333616 | 302365 | 257634 | 227070 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| 1 | 8795 | 99841 | 24816 | 21362 | 22072 | 11629 | 13288 | 27162 | 4688 | 15366 | 15486 |
| 2 | 117383 | 32308 | 127774 | 55025 | 36084 | 53783 | 23145 | 31472 | 54171 | 24969 | 62650 |
| 3 | 18888 | 33973 | 9761 | 43712 | 18056 | 11795 | 16554 | 8523 | 11134 | 20885 | 12753 |
| 4 | 7779 | 5791 | 8689 | 3117 | 9791 | 4299 | 3267 | 4916 | 3126 | 3045 | 5223 |
| 5 | 1369 | 2981 | 1528 | 2543 | 994 | 2445 | 1372 | 1041 | 1546 | 859 | 790 |
| 6 | 1257 | 602 | 1071 | 652 | 1028 | 307 | 1039 | 482 | 426 | 513 | 282 |
| 7 | 371 | 554 | 234 | 293 | 249 | 307 | 222 | 323 | 200 | 140 | 148 |
| 8 | 172 | 170 | 215 | 66 | 139 | 54 | 137 | 51 | 106 | 57 | 41 |
| 9 | 78 | 69 | 55 | 63 | 27 | 60 | 27 | 39 | 17 | 32 | 14 |
| 10 | 16 | 44 | 48 | 23 | 31 | 12 | 4 | 17 | 10 | 7 | 13 |
| +gp | 31 | 23 | 12 | 18 | 10 | 9 | 9 | 9 | 13 | 16 | 5 |
| TOTALNUM | 156139 | 176355 | 174203 | 126873 | 88481 | 84698 | 59065 | 74034 | 75437 | 65889 | 97405 |
| TONSLAND | 214354 | 201279 | 216041 | 183202 | 139578 | 124835 | 101442 | 112740 | 119947 | 109915 | 136397 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 4871 | 23443 | 1243 | 5831 | 8087 | 2164 | 4425 | 438 | 1470 | 1009 | 1286 |
| 2 | 36303 | 28793 | 80948 | 9549 | 22457 | 20309 | 8029 | 8893 | 3511 | 8175 | 4401 |
| 3 | 23046 | 18390 | 16794 | 31624 | 6310 | 6044 | 13831 | 3552 | 5453 | 3036 | 4410 |
| 4 | 3125 | 6409 | 5909 | 3959 | 6529 | 1114 | 2787 | 3072 | 1527 | 1714 | 969 |
| 5 | 1834 | 1221 | 2379 | 1419 | 996 | 1053 | 395 | 397 | 939 | 479 | 520 |
| 6 | 393 | 690 | 504 | 614 | 375 | 140 | 384 | 68 | 155 | 339 | 187 |
| 7 | 159 | 151 | 233 | 219 | 135 | 82 | 58 | 61 | 29 | 52 | 120 |
| 8 | 87 | 47 | 41 | 89 | 39 | 27 | 38 | 15 | 19 | 13 | 23 |
| 9 | 42 | 14 | 16 | 14 | 18 | 13 | 18 | 5 | 6 | 9 | 4 |
| 10 | 4 | 15 | 4 | 10 | 5 | 6 | 4 | 2 | 2 | 1 | 1 |
| +gp | 8 | 10 | 12 | 2 | 1 | 1 | 1 | 0 | 0 | 1 | 0 |
| TOTALNUM | 69872 | 79183 | 108083 | 53329 | 44952 | 30953 | 29971 | 16505 | 13111 | 14830 | 11921 |
| TONSLAND | 124721 | 122434 | 144637 | 94108 | 69567 | 48440 | 53152 | 30426 | 27748 | 28165 | 25665 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |  |  |
| 1 | 776 | 338 | 519 | 1120 | 1099 | 665 | 683 | 2240 | 706 |  |  |
| 2 | 6334 | 3268 | 4833 | 5037 | 4540 | 2230 | 2688 | 4207 | 6430 |  |  |
| 3 | 2264 | 4130 | 2839 | 4578 | 4046 | 5367 | 3063 | 4376 | 4892 |  |  |
| 4 | 1562 | 1146 | 2888 | 1582 | 1408 | 1963 | 2592 | 1605 | 1939 |  |  |
| 5 | 398 | 706 | 596 | 1315 | 610 | 633 | 865 | 1286 | 744 |  |  |
| 6 | 137 | 213 | 237 | 198 | 451 | 248 | 190 | 332 | 583 |  |  |
| 7 | 40 | 70 | 44 | 65 | 48 | 139 | 84 | 64 | 144 |  |  |
| 8 | 39 | 26 | 19 | 16 | 27 | 15 | 38 | 38 | 22 |  |  |
| 9 | 6 | 13 | 17 | 6 | 5 | 4 | 5 | 6 | 6 |  |  |
| 10 | 1 | 1 | 8 | 4 | 2 | 4 | 1 | 2 | 1 |  |  |
| +gp | 1 | 1 | 3 | 2 | 2 | 1 | 1 | 0 | 1 |  |  |
| TOTALNUM | 11558 | 9911 | 12003 | 13923 | 12237 | 11269 | 10208 | 14156 | 15470 |  |  |
| TONSLAND | 24215 | 26814 | 33177 | 36762 | 31979 | 32124 | 30474 | 34651 | 37373 |  |  |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |  |  |

Table 14.2b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Discard numbers at age (Thousands).

| Discards numbers at age (thousands) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE/YEAR | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 |
| 1 | 16150 | 8049 | 97921 | 108375 | 50214 | 31115 | 2502 | 52958 | 258920 | 38250 | 85915 |
| 2 | 19902 | 6168 | 6599 | 22125 | 24736 | 22957 | 10279 | 8656 | 37224 | 59342 | 17387 |
| 3 | 33 | 115 | 89 | 71 | 160 | 197 | 113 | 152 | 47 | 177 | 246 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| +gp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTALNUM | 36085 | 14332 | 104609 | 130570 | 75110 | 54268 | 12894 | 61766 | 296192 | 97768 | 103548 |
| TONSDISC | 12198.57 | 4655.611 | 28972.64 | 37861.71 | 23284.92 | 17468.34 | 4756.776 | 17662.66 | 84006.59 | 33602.62 | 29965.76 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| 1 | 124151 | 136651 | 226781 | 472599 | 28908 | 581071 | 1185689 | 155732 | 181946 | 54949 | 537521 |
| 2 | 15878 | 16214 | 83210 | 48009 | 78114 | 5270 | 17692 | 34307 | 8377 | 11130 | 12518 |
| 3 | 71 | 0 | 192 | 464 | 0 | 0 | 0 | 79 | 98 | 25 | 5 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| +gp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTALNUM | 140100 | 152866 | 310182 | 521072 | 107022 | 586341 | 1203381 | 190118 | 190421 | 66103 | 550043 |
| TONSDISC | 39532.68 | 36840.85 | 72396.83 | 139026.6 | 32433.69 | 162278.1 | 294208.1 | 57075.62 | 54007.83 | 21430.4 | 151003.9 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| 1 | 63301 | 563506 | 24634 | 15376 | 176920 | 33875 | 47473 | 102410 | 33433 | 320725 | 44756 |
| 2 | 36573 | 5761 | 61948 | 17084 | 8685 | 48244 | 8383 | 9881 | 28538 | 16804 | 43434 |
| 3 | 115 | 303 | 0 | 216 | 489 | 78 | 448 | 2 | 11 | 160 | 30 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| +gp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTALNUM | 99989 | 569571 | 86583 | 32676 | 186094 | 82197 | 56304 | 112293 | 61983 | 337689 | 88220 |
| TONSDISC | 31297.6 | 138603.8 | 27706.11 | 10504.47 | 61655.63 | 26747.11 | 18198.97 | 36192.59 | 21411.61 | 98208.27 | 31706.81 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 14254 | 86109 | 15458 | 30962 | 37031 | 5460 | 26267 | 5696 | 20336 | 10213 | 26890 |
| 2 | 23058 | 13701 | 90259 | 5630 | 5509 | 33094 | 13236 | 6082 | 8941 | 8303 | 35342 |
| 3 | 764 | 40 | 1500 | 8280 | 0 | 753 | 3181 | 775 | 2007 | 1795 | 1965 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 17 | 55 | 122 | 149 | 51 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 66 | 4 |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 | 1 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| +gp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTALNUM | 38075 | 99851 | 107216 | 44872 | 42540 | 39307 | 42702 | 12608 | 31413 | 20540 | 64253 |
| TONSDISC | 14030 | 33183.67 | 40102.32 | 13641.52 | 13359.94 | 13519.42 | 11900.56 | 4007.44 | 8721.211 | 9931.799 | 11923 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |  |  |
| 1 | 16171 | 10847 | 9608 | 9867 | 3936 | 11149 | 6188 | 7756 | 4271 |  |  |
| 2 | 23047 | 9331 | 9055 | 9151 | 7851 | 5190 | 6055 | 6504 | 8988 |  |  |
| 3 | 2657 | 7591 | 2655 | 1254 | 925 | 1422 | 856 | 1434 | 1960 |  |  |
| 4 | 481 | 223 | 650 | 65 | 81 | 115 | 397 | 163 | 179 |  |  |
| 5 | 52 | 14 | 50 | 30 | 6 | 5 | 83 | 58 | 54 |  |  |
| 6 | 24 | 11 | 17 | 0 | 4 | 1 | 40 | 5 | 63 |  |  |
| 7 | 0 | 0 | 9 | 0 | 1 | 1 | 16 | 0 | 15 |  |  |
| 8 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 5 |  |  |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |  |  |
| 10 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| +gp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| TOTALNUM | 42433 | 28017 | 22047 | 20366 | 12804 | 17884 | 13635 | 15921 | 15538 |  |  |
| TONSDISC | 30422 | 24984 | 20846 | 12341 | 8711 | 8638 | 10289 | 10538 | 12609 |  |  |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |  |  |

Table 14.2c Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Catch numbers at age (Thousands).

| Catch numbers at age (thousands) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE/YEAR | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 |
| 1 | 19347 | 13052 | 113655 | 126508 | 60962 | 36915 | 5434 | 107177 | 303519 | 42062 | 111751 |
| 2 | 62280 | 28541 | 58227 | 84327 | 95275 | 106373 | 32840 | 42403 | 191789 | 246086 | 48983 |
| 3 | 7028 | 20118 | 17646 | 29766 | 32689 | 42569 | 31532 | 18547 | 17179 | 48062 | 54901 |
| 4 | 3519 | 4285 | 9135 | 6153 | 11205 | 12330 | 13641 | 13272 | 6720 | 5653 | 14002 |
| 5 | 2774 | 1908 | 2375 | 3362 | 3255 | 6046 | 4542 | 6266 | 7065 | 2713 | 2195 |
| 6 | 1207 | 1809 | 946 | 1272 | 1964 | 1407 | 2881 | 1754 | 2686 | 3184 | 1103 |
| 7 | 81 | 596 | 655 | 475 | 884 | 866 | 585 | 956 | 888 | 1671 | 1055 |
| 8 | 489 | 117 | 297 | 368 | 353 | 307 | 420 | 208 | 455 | 609 | 487 |
| 9 | 13 | 93 | 51 | 125 | 137 | 150 | 147 | 185 | 227 | 388 | 79 |
| 10 | 6 | 11 | 75 | 56 | 40 | 111 | 46 | 97 | 77 | 112 | 57 |
| +gp | 0 | 4 | 8 | 83 | 17 | 24 | 77 | 40 | 93 | 17 | 161 |
| TOTALNUM | 96744 | 70535 | 203069 | 252494 | 206780 | 207098 | 92145 | 190905 | 530700 | 350558 | 234774 |
| TONSLAND | 128092 | 130049 | 209092 | 258059 | 274972 | 304417 | 204503 | 242656 | 410498 | 385764 | 267840 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| 1 | 139635 | 169862 | 232476 | 547729 | 58501 | 615698 | 1248084 | 175863 | 248166 | 80437 | 601879 |
| 2 | 74502 | 63121 | 182989 | 98935 | 253025 | 96413 | 122048 | 221933 | 73132 | 139526 | 78543 |
| 3 | 11418 | 18849 | 18672 | 25989 | 17178 | 44384 | 34938 | 34646 | 60005 | 21480 | 31092 |
| 4 | 15745 | 4640 | 6707 | 4597 | 9396 | 4011 | 12274 | 8953 | 9487 | 11787 | 4238 |
| 5 | 4601 | 7525 | 1732 | 2286 | 2989 | 3375 | 1958 | 4088 | 3447 | 2803 | 3415 |
| 6 | 956 | 2057 | 3056 | 833 | 1103 | 708 | 1269 | 779 | 2048 | 1246 | 1013 |
| 7 | 436 | 447 | 920 | 1140 | 408 | 396 | 494 | 599 | 425 | 589 | 434 |
| 8 | 393 | 195 | 130 | 370 | 403 | 139 | 197 | 133 | 234 | 179 | 243 |
| 9 | 330 | 228 | 67 | 262 | 152 | 157 | 73 | 64 | 77 | 89 | 59 |
| 10 | 80 | 95 | 63 | 26 | 36 | 42 | 55 | 36 | 27 | 28 | 44 |
| +gp | 188 | 63 | 43 | 96 | 44 | 17 | 25 | 21 | 16 | 23 | 19 |
| TOTALNUM | 248283 | 267081 | 446854 | 682263 | 343235 | 765338 | 1421415 | 447116 | 397064 | 258186 | 720980 |
| TONSLAND | 252748 | 241089 | 305404 | 347345 | 327074 | 428297 | 587962 | 390691 | 356372 | 279065 | 378074 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| 1 | 72096 | 663347 | 49451 | 36738 | 198992 | 45504 | 60761 | 129572 | 38121 | 336092 | 60242 |
| 2 | 153957 | 38069 | 189722 | 72109 | 44768 | 102027 | 31528 | 41353 | 82709 | 41773 | 106084 |
| 3 | 19003 | 34277 | 9761 | 43929 | 18544 | 11873 | 17002 | 8525 | 11145 | 21045 | 12783 |
| 4 | 7779 | 5791 | 8689 | 3117 | 9791 | 4299 | 3267 | 4916 | 3126 | 3045 | 5223 |
| 5 | 1369 | 2981 | 1528 | 2543 | 994 | 2445 | 1372 | 1041 | 1546 | 859 | 790 |
| 6 | 1257 | 602 | 1071 | 652 | 1028 | 307 | 1039 | 482 | 426 | 513 | 282 |
| 7 | 371 | 554 | 234 | 293 | 249 | 307 | 222 | 323 | 200 | 140 | 148 |
| 8 | 172 | 170 | 215 | 66 | 139 | 54 | 137 | 51 | 106 | 57 | 41 |
| 9 | 78 | 69 | 55 | 63 | 27 | 60 | 27 | 39 | 17 | 32 | 14 |
| 10 | 16 | 44 | 48 | 23 | 31 | 12 | 4 | 17 | 10 | 7 | 13 |
| +gp | 31 | 23 | 12 | 18 | 10 | 9 | 9 | 9 | 13 | 16 | 5 |
| TOTALNUM | 256129 | 745925 | 260786 | 159550 | 274574 | 166895 | 115368 | 186327 | 137419 | 403578 | 185625 |
| TONSLAND | 245651 | 339883 | 243747 | 193706 | 201233 | 151582 | 119641 | 148932 | 141358 | 208123 | 168104 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 19124 | 109552 | 16701 | 36793 | 45118 | 7624 | 30692 | 6135 | 21807 | 11222 | 28177 |
| 2 | 59360 | 42494 | 171206 | 15180 | 27965 | 53403 | 21265 | 14975 | 12452 | 16478 | 39743 |
| 3 | 23809 | 18430 | 18293 | 39904 | 6310 | 6797 | 17012 | 4328 | 7460 | 4831 | 6375 |
| 4 | 3125 | 6409 | 5909 | 3959 | 6529 | 1114 | 2805 | 3127 | 1650 | 1863 | 1020 |
| 5 | 1834 | 1221 | 2379 | 1419 | 996 | 1053 | 395 | 397 | 944 | 546 | 524 |
| 6 | 393 | 690 | 504 | 614 | 375 | 140 | 384 | 68 | 155 | 351 | 187 |
| 7 | 159 | 151 | 233 | 219 | 135 | 82 | 58 | 61 | 29 | 52 | 121 |
| 8 | 87 | 47 | 41 | 89 | 39 | 27 | 38 | 15 | 19 | 13 | 23 |
| 9 | 42 | 14 | 16 | 14 | 18 | 13 | 18 | 5 | 6 | 11 | 4 |
| 10 | 4 | 15 | 4 | 10 | 5 | 6 | 4 | 2 | 2 | 1 | 1 |
| +gp | 8 | 10 | 12 | 2 | 1 | 1 | 1 | 0 | 0 | 1 | 0 |
| TOTALNUM | 107947 | 179034 | 215299 | 98201 | 87491 | 70260 | 72673 | 29113 | 44524 | 35370 | 76174 |
| TONSLAND | 138751 | 155618 | 184740 | 107749 | 82927 | 61960 | 65053 | 34433 | 36469 | 38097 | 37589 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE/YEAR | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |  |  |
| 1 | 16947 | 11185 | 10127 | 10987 | 5035 | 11815 | 6871 | 9995 | 4977 |  |  |
| 2 | 29381 | 12599 | 13887 | 14188 | 12391 | 7420 | 8743 | 10711 | 15418 |  |  |
| 3 | 4921 | 11721 | 5494 | 5831 | 4970 | 6789 | 3919 | 5810 | 6853 |  |  |
| 4 | 2043 | 1369 | 3539 | 1646 | 1489 | 2077 | 2989 | 1768 | 2118 |  |  |
| 5 | 451 | 720 | 646 | 1344 | 616 | 638 | 949 | 1345 | 799 |  |  |
| 6 | 161 | 224 | 254 | 199 | 455 | 249 | 229 | 337 | 647 |  |  |
| 7 | 40 | 70 | 53 | 65 | 49 | 139 | 100 | 64 | 159 |  |  |
| 8 | 41 | 26 | 19 | 16 | 28 | 15 | 38 | 38 | 27 |  |  |
| 9 | 6 | 13 | 17 | 6 | 5 | 4 | 5 | 6 | 9 |  |  |
| 10 | 1 | 1 | 10 | 4 | 2 | 4 | 2 | 2 | 1 |  |  |
| +gp | 1 | 1 | 3 | 2 | 2 | 1 | 1 | 0 | 1 |  |  |
| TOTALNUM | 53992 | 37928 | 34050 | 34288 | 25041 | 29153 | 23844 | 30076 | 31008 |  |  |
| TONSLAND | 54637 | 51798 | 54023 | 49103 | 40689 | 40762 | 40763 | 45190 | 49983 |  |  |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |  |  |

Table 14.2d Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Landings, discards and catch numbers at age (Thousands) by season (quarter or annual, depending on data stratification) from Intercatch for 2015.

| Landings numbers at age (thousands) |  |  |  |  |  |  |
| ---: | :---: | :---: | ---: | ---: | ---: | ---: |
| Age/Season | Q1 | Q2 | Q3 | Q4 | annual | TOTALNUM |
| 1 | 40 | 34 | 125 | 496 | 11 | 706 |
| 2 | 1287 | 1110 | 2081 | 1341 | 611 | 6430 |
| 3 | 1004 | 1067 | 1130 | 1041 | 651 | 4893 |
| 4 | 427 | 502 | 437 | 386 | 187 | 1939 |
| 5 | 188 | 207 | 182 | 118 | 48 | 743 |
| 6 | 154 | 161 | 121 | 73 | 75 | 584 |
| 7 | 30 | 50 | 31 | 17 | 16 | 144 |
| 8 | 6 | 5 | 4 | 2 | 5 | 22 |
| 9 | 1 | 1 | 1 | 1 | 3 | 7 |
| 10 | 0 | 1 | 0 | 0 | 0 | 1 |
| $+g p$ | 0 | 0 | 1 | 1 | 0 | 2 |
| TOTALNUM | 3137 | 3138 | 4113 | 3476 | 1607 | 15471 |


| Discards numbers at age (thousands) |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/Season | Q1 | Q2 | Q3 | Q4 | annual | TOTALNUM |
| 1 | 470 | 372 | 1542 | 1117 | 770 | 4271 |
| 2 | 1287 | 1336 | 1804 | 1787 | 2774 | 8988 |
| 3 | 359 | 575 | 320 | 344 | 362 | 1960 |
| 4 | 46 | 40 | 34 | 31 | 28 | 179 |
| 5 | 31 | 7 | 4 | 4 | 8 | 54 |
| 6 | 39 | 7 | 6 | 4 | 7 | 63 |
| 7 | 10 | 2 | 1 | 1 | 0 | 14 |
| 8 | 4 | 0 | 0 | 0 | 0 | 4 |
| 9 | 1 | 0 | 1 | 0 | 0 | 2 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 |
| $+g p$ | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTALNUM | 2247 | 2339 | 3712 | 3288 | 3949 | 15535 |


| Catch numbers at age (thousands) |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/Season | Q1 | Q2 | Q3 | Q4 | annual | TOTALNUM |
| 1 | 510 | 405 | 1667 | 1614 | 781 | 4977 |
| 2 | 2574 | 2446 | 3885 | 3128 | 3385 | 15418 |
| 3 | 1363 | 1642 | 1451 | 1385 | 1013 | 6854 |
| 4 | 473 | 543 | 470 | 416 | 215 | 2117 |
| 5 | 219 | 215 | 187 | 123 | 56 | 800 |
| 6 | 194 | 168 | 127 | 77 | 82 | 648 |
| 7 | 40 | 52 | 32 | 18 | 17 | 159 |
| 8 | 10 | 5 | 4 | 3 | 5 | 27 |
| 9 | 2 | 1 | 2 | 1 | 3 | 9 |
| 10 | 0 | 1 | 0 | 0 | 0 | 1 |
| $+g p$ | 0 | 0 | 1 | 1 | 0 | 2 |
| TOTALNUM | 5385 | 5478 | 7826 | 6766 | 5557 | 31012 |

Table 14.2e Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Sampling coverage for discard ratio, landings age composition and discards age composition by area and season (quarter or annual, depending on data stratification) for 2015, calculated as the weight in each area-season-métier stratum covered by the relevant sampling, then summed over métiers and expressed as a proportion of the total for the area-season (note the country dimension is not used). Also provided is the contribution of landings and discards in each area (by weight) to the total for that catch category (before raising is conducted).

Discard ratio coverage

| Area/Season | Q1 | Q2 | Q3 | Q4 | annual |
| :--- | ---: | ---: | ---: | ---: | ---: |
| IV | $75 \%$ | $66 \%$ | $65 \%$ | $69 \%$ | $96 \%$ |
| IllaN | $78 \%$ | $75 \%$ | $57 \%$ | $87 \%$ | - |
| VIId | $84 \%$ | $89 \%$ | $84 \%$ | $77 \%$ | - |

Landings age composition coverage

| Area/Season | Q1 | Q2 | Q3 | Q4 | annual |
| :--- | ---: | ---: | ---: | ---: | ---: |
| IV | $81 \%$ | $78 \%$ | $75 \%$ | $77 \%$ | $96 \%$ |
| IIIaN | $94 \%$ | $95 \%$ | $77 \%$ | $90 \%$ | - |
| VIId | $84 \%$ | $87 \%$ | $87 \%$ | $87 \%$ | - |

Discards age composition coverage

| Area/Season | Q1 | Q2 | Q3 | Q4 | annual |
| :--- | ---: | ---: | ---: | ---: | ---: |
| IV | $98 \%$ | $90 \%$ | $96 \%$ | $93 \%$ | $100 \%$ |
| IIlaN | $99 \%$ | $100 \%$ | $100 \%$ | $92 \%$ | - |
| VIId | $100 \%$ | $100 \%$ | $100 \%$ | $100 \%$ | - |

Contribution to total (before raising)

| Area/Type | Landings | Discards |
| :--- | ---: | ---: |
| IV | $84 \%$ | $76 \%$ |
| IllaN | $12 \%$ | $23 \%$ |
| VIId | $4 \%$ | $0 \%$ |

Table 14.3a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Landings weights at age (kg).

| Landings weights at age (kg) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE/YEAR | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 |
| 1 | 0.538 | 0.496 | 0.581 | 0.579 | 0.590 | 0.640 | 0.544 | 0.626 | 0.579 | 0.616 | 0.559 |
| 2 | 1.004 | 0.863 | 0.965 | 0.994 | 1.035 | 0.973 | 0.921 | 0.961 | 0.941 | 0.836 | 0.869 |
| 3 | 2.657 | 2.377 | 2.304 | 2.442 | 2.404 | 2.223 | 2.133 | 2.041 | 2.193 | 2.086 | 1.919 |
| 4 | 4.491 | 4.528 | 4.512 | 4.169 | 3.153 | 4.094 | 3.852 | 4.001 | 4.258 | 3.968 | 3.776 |
| 5 | 6.794 | 6.447 | 7.274 | 7.027 | 6.803 | 5.341 | 5.715 | 6.131 | 6.528 | 6.011 | 5.488 |
| 6 | 9.409 | 8.520 | 9.498 | 9.599 | 9.610 | 8.020 | 6.722 | 7.945 | 8.646 | 8.246 | 7.453 |
| 7 | 11.562 | 10.606 | 11.898 | 11.766 | 12.033 | 8.581 | 9.262 | 9.953 | 10.356 | 9.766 | 9.019 |
| 8 | 11.942 | 10.758 | 12.041 | 11.968 | 12.481 | 10.162 | 9.749 | 10.131 | 11.219 | 10.228 | 9.810 |
| 9 | 13.383 | 12.340 | 13.053 | 14.060 | 13.589 | 10.720 | 10.384 | 11.919 | 12.881 | 11.875 | 11.077 |
| 10 | 13.756 | 12.540 | 14.441 | 14.746 | 14.271 | 12.497 | 12.743 | 12.554 | 13.147 | 12.530 | 12.359 |
| +gp | 0.000 | 18.000 | 15.667 | 15.672 | 19.016 | 11.595 | 11.175 | 14.367 | 15.544 | 14.350 | 12.886 |
| AGE/YEAR | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| 1 | 0.594 | 0.619 | 0.568 | 0.541 | 0.573 | 0.550 | 0.550 | 0.723 | 0.589 | 0.632 | 0.594 |
| 2 | 1.039 | 0.899 | 1.029 | 0.948 | 0.937 | 0.936 | 1.003 | 0.837 | 0.962 | 0.919 | 1.007 |
| 3 | 2.217 | 2.348 | 2.470 | 2.160 | 2.001 | 2.411 | 1.948 | 2.190 | 1.858 | 1.835 | 2.156 |
| 4 | 4.156 | 4.226 | 4.577 | 4.606 | 4.146 | 4.423 | 4.401 | 4.615 | 4.130 | 3.880 | 3.972 |
| 5 | 6.174 | 6.404 | 6.494 | 6.714 | 6.530 | 6.579 | 6.109 | 7.045 | 6.785 | 6.491 | 6.190 |
| 6 | 8.333 | 8.691 | 8.620 | 8.828 | 8.667 | 8.474 | 9.120 | 8.884 | 8.903 | 8.423 | 8.362 |
| 7 | 9.889 | 10.107 | 10.132 | 10.071 | 9.685 | 10.637 | 9.550 | 9.933 | 10.398 | 9.848 | 10.317 |
| 8 | 10.791 | 10.910 | 11.340 | 11.052 | 11.099 | 11.550 | 11.867 | 11.519 | 12.500 | 11.837 | 11.352 |
| 9 | 12.175 | 12.339 | 12.888 | 11.824 | 12.427 | 13.057 | 12.782 | 13.338 | 13.469 | 12.797 | 13.505 |
| 10 | 12.425 | 12.976 | 14.139 | 13.134 | 12.778 | 14.148 | 14.081 | 14.897 | 12.890 | 12.562 | 13.408 |
| +gp | 13.731 | 14.431 | 14.760 | 14.362 | 13.981 | 15.478 | 15.392 | 18.784 | 14.608 | 14.426 | 13.472 |
| AGE/YEAR | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| 1 | 0.590 | 0.583 | 0.635 | 0.585 | 0.673 | 0.737 | 0.670 | 0.699 | 0.699 | 0.677 | 0.721 |
| 2 | 0.932 | 0.856 | 0.976 | 0.881 | 1.052 | 0.976 | 1.078 | 1.146 | 1.065 | 1.075 | 1.021 |
| 3 | 2.141 | 1.834 | 1.955 | 1.982 | 1.846 | 2.176 | 2.038 | 2.546 | 2.479 | 2.201 | 2.210 |
| 4 | 4.164 | 3.504 | 3.650 | 3.187 | 3.585 | 3.791 | 3.971 | 4.223 | 4.551 | 4.471 | 4.293 |
| 5 | 6.324 | 6.230 | 6.052 | 5.992 | 5.273 | 5.931 | 6.082 | 6.247 | 6.540 | 7.167 | 7.220 |
| 6 | 8.430 | 8.140 | 8.307 | 7.914 | 7.921 | 7.890 | 8.033 | 8.483 | 8.094 | 8.436 | 8.980 |
| 7 | 10.362 | 9.896 | 10.243 | 9.764 | 9.724 | 10.235 | 9.545 | 10.101 | 9.641 | 9.537 | 10.282 |
| 8 | 12.074 | 11.940 | 11.461 | 12.127 | 11.212 | 10.923 | 10.948 | 10.482 | 10.734 | 10.323 | 11.743 |
| 9 | 13.072 | 12.951 | 12.447 | 14.242 | 12.586 | 12.803 | 13.481 | 11.849 | 12.329 | 12.223 | 13.107 |
| 10 | 14.443 | 13.859 | 18.691 | 17.787 | 15.557 | 15.525 | 13.171 | 13.904 | 13.443 | 14.247 | 12.052 |
| +gp | 16.588 | 14.707 | 16.604 | 16.477 | 14.695 | 23.234 | 14.989 | 15.794 | 13.961 | 12.523 | 13.954 |
| AGE/YEAR | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 0.699 | 0.656 | 0.542 | 0.640 | 0.611 | 0.725 | 0.626 | 0.573 | 0.726 | 0.747 | 0.793 |
| 2 | 1.117 | 0.960 | 0.922 | 0.935 | 1.021 | 1.004 | 0.996 | 1.079 | 1.072 | 1.160 | 1.200 |
| 3 | 2.147 | 2.120 | 1.724 | 1.663 | 1.747 | 2.303 | 1.844 | 1.895 | 2.089 | 1.952 | 2.239 |
| 4 | 4.034 | 3.821 | 3.495 | 3.305 | 3.216 | 3.663 | 3.735 | 3.347 | 3.252 | 3.647 | 3.894 |
| 5 | 6.637 | 6.228 | 5.387 | 5.726 | 4.903 | 5.871 | 5.537 | 5.757 | 5.184 | 5.244 | 5.676 |
| 6 | 8.494 | 8.394 | 7.563 | 7.403 | 7.488 | 7.333 | 8.006 | 6.694 | 7.438 | 7.225 | 7.234 |
| 7 | 9.729 | 9.979 | 9.628 | 8.582 | 9.636 | 9.264 | 9.451 | 8.838 | 8.974 | 9.457 | 9.243 |
| 8 | 11.080 | 11.424 | 10.643 | 10.365 | 10.671 | 10.081 | 10.012 | 12.674 | 9.894 | 10.567 | 10.477 |
| 9 | 12.264 | 12.300 | 11.499 | 11.600 | 10.894 | 12.062 | 11.888 | 11.518 | 11.857 | 12.015 | 12.325 |
| 10 | 12.756 | 12.761 | 13.085 | 12.330 | 11.414 | 12.009 | 12.795 | 11.053 | 12.095 | 12.066 | 14.862 |
| +gp | 11.304 | 13.416 | 14.921 | 11.926 | 15.078 | 10.196 | 11.688 | 14.988 | 14.093 | 22.464 | 17.887 |
| AGE/YEAR | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |  |  |
| 1 | 0.830 | 1.06679 | 0.78826 | 0.71481 | 0.862 | 0.938 | 0.883 | 0.699 | 0.594 |  |  |
| 2 | 1.182 | 1.38884 | 1.41193 | 1.29224 | 1.328 | 1.369 | 1.240 | 1.213 | 1.198 |  |  |
| 3 | 2.365 | 2.45605 | 2.67433 | 2.67091 | 2.525 | 2.354 | 2.461 | 2.390 | 2.290 |  |  |
| 4 | 4.050 | 4.06299 | 4.14457 | 4.22308 | 4.596 | 4.175 | 4.164 | 4.180 | 4.111 |  |  |
| 5 | 6.053 | 6.22405 | 6.11913 | 6.04897 | 6.481 | 6.391 | 6.187 | 5.678 | 5.935 |  |  |
| 6 | 8.250 | 7.39317 | 7.48963 | 8.29925 | 7.843 | 8.115 | 8.347 | 7.435 | 6.923 |  |  |
| 7 | 9.262 | 9.65076 | 8.96797 | 9.47215 | 9.681 | 9.092 | 9.817 | 9.191 | 8.774 |  |  |
| 8 | 10.015 | 11.48868 | 11.44744 | 11.63072 | 9.629 | 11.799 | 9.486 | 9.180 | 9.627 |  |  |
| 9 | 12.282 | 11.38721 | 11.29135 | 12.82728 | 10.845 | 12.548 | 11.364 | 11.469 | 10.654 |  |  |
| 10 | 14.559 | 12.72507 | 11.71648 | 12.08332 | 14.436 | 11.436 | 10.935 | 16.456 | 13.838 |  |  |
| +gp | 17.522 | 15.38134 | 18.764 | 10.05238 | 12.421 | 20.644 | 29.764 | 34.656 | 30.079 |  |  |

Table 14.3b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Discard weights at age (kg).

| Discards weights at age (kg) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE/YEAI | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 |
| 1 | 0.270 | 0.270 | 0.269 | 0.269 | 0.269 | 0.269 | 0.268 | 0.268 | 0.268 | 0.268 | 0.268 |
| 2 | 0.393 | 0.393 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 |
| 3 | 0.505 | 0.508 | 0.506 | 0.509 | 0.506 | 0.505 | 0.504 | 0.505 | 0.508 | 0.507 | 0.507 |
| 4 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 6 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 7 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| +gp | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| AGE/YEAI | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| 1 | 0.268 | 0.227 | 0.189 | 0.255 | 0.287 | 0.276 | 0.242 | 0.279 | 0.274 | 0.297 | 0.270 |
| 2 | 0.392 | 0.359 | 0.354 | 0.382 | 0.309 | 0.361 | 0.411 | 0.396 | 0.489 | 0.458 | 0.469 |
| 3 | 0.508 | 0.000 | 0.412 | 0.376 | 0.000 | 0.000 | 0.000 | 0.517 | 0.593 | 0.534 | 0.509 |
| 4 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 6 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 7 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| +gp | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| AGE/YEAI | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| 1 | 0.276 | 0.242 | 0.237 | 0.300 | 0.326 | 0.260 | 0.315 | 0.314 | 0.274 | 0.287 | 0.316 |
| 2 | 0.376 | 0.365 | 0.353 | 0.339 | 0.431 | 0.371 | 0.366 | 0.408 | 0.429 | 0.362 | 0.404 |
| 3 | 0.652 | 0.437 | 0.000 | 0.463 | 0.484 | 0.526 | 0.395 | 2.309 | 0.705 | 0.483 | 0.553 |
| 4 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 6 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 7 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| +gp | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| AGE/YEAI | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 0.342 | 0.313 | 0.358 | 0.257 | 0.298 | 0.232 | 0.243 | 0.262 | 0.236 | 0.302 | 0.224 |
| 2 | 0.380 | 0.453 | 0.375 | 0.389 | 0.422 | 0.361 | 0.314 | 0.345 | 0.270 | 0.565 | 0.116 |
| 3 | 0.515 | 0.616 | 0.481 | 0.422 | 0.000 | 0.406 | 0.413 | 0.498 | 0.686 | 0.814 | 0.827 |
| 4 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.205 | 0.528 | 0.864 | 2.223 | 2.557 |
| 5 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 3.852 | 4.255 | 4.208 |
| 6 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 11.300 | 6.509 | 5.437 |
| 7 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 11.048 |
| 8 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 8.100 | 0.000 |
| 10 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| +gp | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| AGE/YEAI | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |  |  |
| 1 | 0.288 | 0.404 | 0.385 | 0.292 | 0.277 | 0.234 | 0.334 | 0.311 | 0.321 |  |  |
| 2 | 0.814 | 0.735 | 0.984 | 0.785 | 0.677 | 0.556 | 0.796 | 0.742 | 0.756 |  |  |
| 3 | 1.690 | 1.699 | 2.013 | 1.533 | 2.057 | 1.867 | 1.493 | 1.772 | 1.616 |  |  |
| 4 | 3.949 | 3.002 | 3.485 | 3.137 | 4.099 | 3.803 | 3.375 | 3.128 | 3.157 |  |  |
| 5 | 6.609 | 5.311 | 6.565 | 5.323 | 5.576 | 6.456 | 4.048 | 3.826 | 3.983 |  |  |
| 6 | 10.198 | 9.341 | 8.521 | 8.369 | 6.071 | 8.579 | 8.419 | 4.642 | 5.303 |  |  |
| 7 | 5.900 | 5.128 | 13.464 | 6.728 | 8.264 | 9.733 | 7.086 | 4.423 | 6.940 |  |  |
| 8 | 15.906 | 0.000 | 0.000 | 0.000 | 6.213 | 0.000 | 0.000 | 0.000 | 8.390 |  |  |
| 9 | 0.000 | 0.000 | 0.000 | 0.000 | 11.617 | 0.000 | 0.000 | 0.000 | 4.081 |  |  |
| 10 | 0.000 | 0.000 | 12.014 | 0.000 | 0.000 | 16.370 | 16.370 | 0.000 | 0.000 |  |  |
| +gp | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |  |  |

Table 14.3c Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Catch weights at age (kg), also assumed to represent stock weights at age.

| Catch weights at age (kg) |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AGE/YEAI | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 |
| 1 | 0.314 | 0.357 | 0.312 | 0.313 | 0.326 | 0.327 | 0.417 | 0.449 | 0.314 | 0.300 |
| 2 | 0.809 | 0.761 | 0.900 | 0.836 | 0.868 | 0.848 | 0.755 | 0.845 | 0.834 | 0.729 |
| 3 | 2.647 | 2.366 | 2.295 | 2.437 | 2.395 | 2.215 | 2.127 | 2.028 | 2.188 | 2.080 |
| 4 | 4.491 | 4.528 | 4.512 | 4.169 | 3.153 | 4.094 | 3.852 | 4.001 | 4.258 | 3.968 |
| 5 | 6.794 | 6.447 | 7.274 | 7.027 | 6.803 | 5.341 | 5.715 | 6.131 | 6.528 | 6.011 |
| 6 | 9.409 | 8.520 | 9.498 | 9.599 | 9.610 | 8.020 | 6.722 | 7.945 | 8.646 | 8.246 |
| 7 | 11.562 | 10.606 | 11.898 | 11.766 | 12.033 | 8.581 | 9.262 | 9.953 | 10.356 | 9.766 |
| 8 | 11.942 | 10.758 | 12.041 | 11.968 | 12.481 | 10.162 | 9.749 | 10.131 | 11.219 | 10.228 |
| 9 | 13.383 | 12.340 | 13.053 | 14.060 | 13.589 | 10.720 | 10.384 | 11.919 | 12.881 | 11.875 |
| 10 | 13.756 | 12.540 | 14.441 | 14.746 | 14.271 | 12.497 | 12.743 | 12.554 | 13.147 | 12.530 |

Table 14.3d Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Landings, discards and catch weights at age (kg) by season (quarter or annual, depending on data stratification) from Intercatch for 2015 (note, any differences in the +gp values between Tables 14.3a-c and Table 14.3d is due to rounding error alone).

Landings weights at age (kg)

| Age/Season | Q1 | Q2 | Q3 | Q4 | annual | total |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.555 | 0.595 | 0.593 | 0.593 | 0.803 | 0.594 |
| 2 | 1.009 | 1.12 | 1.235 | 1.427 | 1.115 | 1.198 |
| 3 | 1.903 | 2.004 | 2.596 | 2.867 | 1.901 | 2.29 |
| 4 | 3.421 | 3.782 | 4.71 | 4.849 | 3.655 | 4.111 |
| 5 | 5.159 | 5.784 | 6.497 | 6.673 | 5.671 | 5.935 |
| 6 | 6.266 | 6.936 | 7.685 | 7.842 | 6.131 | 6.923 |
| 7 | 8.889 | 8.315 | 9.464 | 9.765 | 7.607 | 8.774 |
| 8 | 8.751 | 10.901 | 9.762 | 11.634 | 8.379 | 9.627 |
| 9 | 12.07 | 11.699 | 11.276 | 11.994 | 9.336 | 10.654 |
| 10 | 12.631 | 12.838 | 15.512 | 15.483 | 12.665 | 13.838 |
| +gp | 33.87 | 23.635 | 26.772 | 37.213 | 30.642 | 30.079 |

Discards weights at age (kg)

| Age/Season | Q 1 | Q 2 | Q 3 | Q 4 | annual | total |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.215 | 0.286 | 0.33 | 0.412 | 0.254 | 0.321 |
| 2 | 0.592 | 0.687 | 0.842 | 1.061 | 0.615 | 0.756 |
| 3 | 1.345 | 1.451 | 1.791 | 2.09 | 1.541 | 1.616 |
| 4 | 3.269 | 3.193 | 3.039 | 3.162 | 3.061 | 3.157 |
| 5 | 3.823 | 4.597 | 4.033 | 4.386 | 3.765 | 3.983 |
| 6 | 5.733 | 4.466 | 5.553 | 5.502 | 3.405 | 5.303 |
| 7 | 7.562 | 4.49 | 6.94 | 6.94 | 6.94 | 6.94 |
| 8 | 8.39 | 8.39 | 8.39 | 8.39 | 8.39 | 8.39 |
| 9 | 7.2 | 4.516 | 0.805 | 4.774 | 6.051 | 4.081 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 |
| $+g p$ | 0 | 0 | 0 | 0 | 0 | 0 |

Catch weights at age (kg)

| Age/Season | Q 1 | Q 2 | Q 3 | Q 4 | annual | total |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.242 | 0.312 | 0.35 | 0.468 | 0.262 | 0.36 |
| 2 | 0.8 | 0.883 | 1.052 | 1.218 | 0.705 | 0.941 |
| 3 | 1.756 | 1.811 | 2.418 | 2.674 | 1.772 | 2.097 |
| 4 | 3.406 | 3.738 | 4.591 | 4.724 | 3.577 | 4.031 |
| 5 | 4.973 | 5.743 | 6.439 | 6.59 | 5.413 | 5.802 |
| 6 | 6.158 | 6.834 | 7.583 | 7.718 | 5.896 | 6.764 |
| 7 | 8.563 | 8.132 | 9.366 | 9.608 | 7.595 | 8.603 |
| 8 | 8.624 | 10.767 | 9.634 | 11.258 | 8.379 | 9.416 |
| 9 | 9.524 | 10.375 | 5.486 | 10.692 | 9.294 | 8.67 |
| 10 | 12.631 | 12.838 | 15.512 | 15.483 | 12.665 | 13.838 |
| $+g p$ | 33.87 | 23.635 | 26.772 | 37.213 | 30.642 | 30.079 |

Table 14.4 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Reported landings, estimated discards and total catch (landings + discards) in tonnes. Note any differences in values between Table 14.4 and those given in the report and advice are due to SOP correction.

| year | landings | discards | catch |
| ---: | ---: | ---: | ---: |
| 1963 | 115893 | 12199 | 128092 |
| 1964 | 125393 | 4656 | 130049 |
| 1965 | 180120 | 28973 | 209092 |
| 1966 | 220197 | 37862 | 258059 |
| 1967 | 251687 | 23285 | 274972 |
| 1968 | 286948 | 17468 | 304417 |
| 1969 | 199746 | 4757 | 204503 |
| 1970 | 224993 | 17663 | 242656 |
| 1971 | 326492 | 84007 | 410498 |
| 1972 | 352161 | 33603 | 385764 |
| 1973 | 237874 | 29966 | 267840 |
| 1974 | 213215 | 39533 | 252748 |
| 1975 | 204249 | 36841 | 241089 |
| 1976 | 233007 | 72397 | 305404 |
| 1977 | 208318 | 139027 | 347345 |
| 1978 | 294640 | 32434 | 327074 |
| 1979 | 266019 | 162278 | 428297 |
| 1980 | 293753 | 294208 | 587962 |
| 1981 | 333616 | 57076 | 390691 |
| 1982 | 302365 | 54008 | 356372 |
| 1983 | 257634 | 21430 | 279065 |
| 1984 | 227070 | 151004 | 378074 |
| 1985 | 214354 | 31298 | 245651 |
| 1986 | 201279 | 138604 | 339883 |
| 1987 | 216041 | 27706 | 243747 |
| 1988 | 183202 | 10504 | 193706 |
| 1989 | 139578 | 61656 | 201233 |
| 1990 | 124835 | 26747 | 151582 |
| 1991 | 101442 | 18199 | 119641 |
| 1992 | 112740 | 36193 | 148932 |
| 1993 | 119947 | 21412 | 141358 |
| 1994 | 109915 | 98208 | 208123 |
| 1995 | 136397 | 31707 | 168104 |
| 1996 | 124721 | 14030 | 138751 |
| 1997 | 122434 | 33184 | 155618 |
| 1998 | 144637 | 40102 | 184740 |
| 1999 | 94108 | 13642 | 107749 |
| 2000 | 69567 | 13360 | 82927 |
| 2001 | 48440 | 13519 | 61960 |
| 2002 | 53152 | 11901 | 65053 |
| 2003 | 30426 | 4007 | 34433 |
| 2004 | 27748 | 8721 | 36469 |
| 2005 | 28165 | 9932 | 38097 |
| 2006 | 25665 | 11923 | 37589 |
| 2007 | 24215 | 30422 | 54637 |
| 2008 | 26814 | 24984 | 51798 |
| 2009 | 33177 | 20846 | 54023 |
| 2010 | 36762 | 12341 | 49103 |
| 2011 | 31979 | 8711 | 40689 |
| 2012 | 32124 | 8638 | 40762 |
| 2013 | 30474 | 10289 | 40763 |
| 2014 | 34651 | 10538 | 45190 |
| 2015 | 37373 | 12609 | 49983 |
|  |  |  |  |
|  |  |  |  |
|  |  |  |  |

Table 14.5a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Proportion mature by agegroup.

|  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6+ |
| 1963 | 0.010 | 0.050 | 0.230 | 0.620 | 0.860 | 1.000 |
| 1964 | 0.010 | 0.050 | 0.230 | 0.620 | 0.860 | 1.000 |
| 1965 | 0.010 | 0.050 | 0.230 | 0.620 | 0.860 | 1.000 |
| 1966 | 0.010 | 0.050 | 0.230 | 0.620 | 0.860 | 1.000 |
| 1967 | 0.010 | 0.050 | 0.230 | 0.620 | 0.860 | 1.000 |
| 1968 | 0.010 | 0.050 | 0.230 | 0.620 | 0.860 | 1.000 |
| 1969 | 0.010 | 0.050 | 0.230 | 0.620 | 0.860 | 1.000 |
| 1970 | 0.010 | 0.050 | 0.230 | 0.620 | 0.860 | 1.000 |
| 1971 | 0.010 | 0.050 | 0.230 | 0.620 | 0.860 | 1.000 |
| 1972 | 0.010 | 0.050 | 0.230 | 0.620 | 0.860 | 1.000 |
| 1973 | 0.008 | 0.030 | 0.228 | 0.628 | 0.859 | 1.000 |
| 1974 | 0.007 | 0.035 | 0.223 | 0.616 | 0.846 | 1.000 |
| 1975 | 0.007 | 0.040 | 0.220 | 0.603 | 0.833 | 1.000 |
| 1976 | 0.006 | 0.046 | 0.217 | 0.591 | 0.820 | 1.000 |
| 1977 | 0.005 | 0.051 | 0.215 | 0.580 | 0.809 | 1.000 |
| 1978 | 0.005 | 0.056 | 0.215 | 0.570 | 0.799 | 1.000 |
| 1979 | 0.004 | 0.061 | 0.216 | 0.560 | 0.791 | 1.000 |
| 1980 | 0.003 | 0.067 | 0.219 | 0.551 | 0.785 | 1.000 |
| 1981 | 0.003 | 0.072 | 0.225 | 0.544 | 0.782 | 1.000 |
| 1982 | 0.002 | 0.078 | 0.233 | 0.539 | 0.780 | 1.000 |
| 1983 | 0.002 | 0.084 | 0.245 | 0.538 | 0.781 | 1.000 |
| 1984 | 0.002 | 0.090 | 0.262 | 0.541 | 0.785 | 1.000 |
| 1985 | 0.001 | 0.096 | 0.283 | 0.549 | 0.791 | 1.000 |
| 1986 | 0.001 | 0.103 | 0.309 | 0.564 | 0.800 | 1.000 |
| 1987 | 0.001 | 0.110 | 0.339 | 0.586 | 0.812 | 1.000 |
| 1988 | 0.001 | 0.117 | 0.371 | 0.612 | 0.825 | 1.000 |
| 1989 | 0.001 | 0.124 | 0.402 | 0.643 | 0.840 | 1.000 |
| 1990 | 0.002 | 0.132 | 0.431 | 0.675 | 0.856 | 1.000 |
| 1991 | 0.002 | 0.141 | 0.454 | 0.706 | 0.872 | 1.000 |
| 1992 | 0.003 | 0.149 | 0.471 | 0.734 | 0.887 | 1.000 |
| 1993 | 0.003 | 0.158 | 0.483 | 0.757 | 0.901 | 1.000 |
| 1994 | 0.004 | 0.167 | 0.489 | 0.774 | 0.913 | 1.000 |
| 1995 | 0.005 | 0.177 | 0.492 | 0.787 | 0.923 | 1.000 |
| 1996 | 0.006 | 0.187 | 0.495 | 0.794 | 0.931 | 1.000 |
| 1997 | 0.008 | 0.197 | 0.501 | 0.798 | 0.937 | 1.000 |
| 1998 | 0.009 | 0.208 | 0.511 | 0.799 | 0.941 | 1.000 |
| 1999 | 0.011 | 0.218 | 0.525 | 0.799 | 0.943 | 1.000 |
| 2000 | 0.013 | 0.230 | 0.545 | 0.798 | 0.944 | 1.000 |
| 2001 | 0.015 | 0.241 | 0.569 | 0.798 | 0.944 | 1.000 |
| 2002 | 0.017 | 0.252 | 0.596 | 0.798 | 0.944 | 1.000 |
| 2003 | 0.019 | 0.264 | 0.624 | 0.800 | 0.943 | 1.000 |
| 2004 | 0.022 | 0.276 | 0.651 | 0.804 | 0.943 | 1.000 |
| 2005 | 0.024 | 0.287 | 0.676 | 0.810 | 0.943 | 1.000 |
| 2006 | 0.027 | 0.299 | 0.697 | 0.818 | 0.944 | 1.000 |
| 2007 | 0.030 | 0.310 | 0.714 | 0.828 | 0.945 | 1.000 |
| 2008 | 0.033 | 0.322 | 0.725 | 0.839 | 0.946 | 1.000 |
| 2009 | 0.036 | 0.333 | 0.729 | 0.850 | 0.947 | 0.990 |
| 2010 | 0.039 | 0.345 | 0.727 | 0.859 | 0.947 | 0.970 |
| 2011 | 0.041 | 0.357 | 0.718 | 0.867 | 0.945 | 1.000 |
| 2012 | 0.044 | 0.368 | 0.702 | 0.872 | 0.943 | 1.000 |
| 2013 | 0.047 | 0.380 | 0.682 | 0.875 | 0.939 | 1.000 |
| 2014 | 0.050 | 0.392 | 0.659 | 0.877 | 0.934 | 1.000 |
| 2015 | 0.053 | 0.403 | 0.634 | 0.878 | 0.929 | 1.000 |

Table 14.5b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Natural mortality by agegroup.

| y | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 |
| 1963 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1964 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1965 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1966 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1967 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1968 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1969 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1970 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1971 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1972 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1973 | 1.215 | 0.777 | 0.221 | 0.2 | 0.2 | 0.2 |
| 1974 | 1.208 | 0.767 | 0.211 | 0.2 | 0.2 | 0.2 |
| 1975 | 1.233 | 0.746 | 0.211 | 0.2 | 0.2 | 0.2 |
| 1976 | 1.260 | 0.729 | 0.211 | 0.2 | 0.2 | 0.2 |
| 1977 | 1.286 | 0.715 | 0.211 | 0.2 | 0.2 | 0.2 |
| 1978 | 1.311 | 0.705 | 0.211 | 0.2 | 0.2 | 0.2 |
| 1979 | 1.332 | 0.701 | 0.211 | 0.2 | 0.2 | 0.2 |
| 1980 | 1.349 | 0.702 | 0.211 | 0.2 | 0.2 | 0.2 |
| 1981 | 1.360 | 0.706 | 0.211 | 0.2 | 0.2 | 0.2 |
| 1982 | 1.362 | 0.710 | 0.211 | 0.2 | 0.2 | 0.2 |
| 1983 | 1.357 | 0.715 | 0.212 | 0.2 | 0.2 | 0.2 |
| 1984 | 1.344 | 0.717 | 0.212 | 0.2 | 0.2 | 0.2 |
| 1985 | 1.325 | 0.718 | 0.213 | 0.2 | 0.2 | 0.2 |
| 1986 | 1.301 | 0.718 | 0.213 | 0.2 | 0.2 | 0.2 |
| 1987 | 1.274 | 0.718 | 0.214 | 0.2 | 0.2 | 0.2 |
| 1988 | 1.247 | 0.718 | 0.215 | 0.2 | 0.2 | 0.2 |
| 1989 | 1.220 | 0.720 | 0.215 | 0.2 | 0.2 | 0.2 |
| 1990 | 1.196 | 0.722 | 0.216 | 0.2 | 0.2 | 0.2 |
| 1991 | 1.174 | 0.723 | 0.216 | 0.2 | 0.2 | 0.2 |
| 1992 | 1.157 | 0.725 | 0.217 | 0.2 | 0.2 | 0.2 |
| 1993 | 1.144 | 0.727 | 0.217 | 0.2 | 0.2 | 0.2 |
| 1994 | 1.136 | 0.730 | 0.217 | 0.2 | 0.2 | 0.2 |
| 1995 | 1.129 | 0.734 | 0.218 | 0.2 | 0.2 | 0.2 |
| 1996 | 1.122 | 0.740 | 0.219 | 0.2 | 0.2 | 0.2 |
| 1997 | 1.115 | 0.748 | 0.220 | 0.2 | 0.2 | 0.2 |
| 1998 | 1.106 | 0.756 | 0.222 | 0.2 | 0.2 | 0.2 |
| 1999 | 1.097 | 0.767 | 0.224 | 0.2 | 0.2 | 0.2 |
| 2000 | 1.088 | 0.779 | 0.226 | 0.2 | 0.2 | 0.2 |
| 2001 | 1.084 | 0.795 | 0.229 | 0.2 | 0.2 | 0.2 |
| 2002 | 1.085 | 0.814 | 0.232 | 0.2 | 0.2 | 0.2 |
| 2003 | 1.091 | 0.835 | 0.235 | 0.2 | 0.2 | 0.2 |
| 2004 | 1.100 | 0.854 | 0.237 | 0.2 | 0.2 | 0.2 |
| 2005 | 1.112 | 0.871 | 0.238 | 0.2 | 0.2 | 0.2 |
| 2006 | 1.126 | 0.884 | 0.239 | 0.2 | 0.2 | 0.2 |
| 2007 | 1.141 | 0.893 | 0.238 | 0.2 | 0.2 | 0.2 |
| 2008 | 1.159 | 0.900 | 0.237 | 0.2 | 0.2 | 0.2 |
| 2009 | 1.180 | 0.907 | 0.236 | 0.2 | 0.2 | 0.2 |
| 2010 | 1.208 | 0.916 | 0.235 | 0.2 | 0.2 | 0.2 |
| 2011 | 1.242 | 0.929 | 0.234 | 0.2 | 0.2 | 0.2 |
| 2012 | 1.283 | 0.945 | 0.233 | 0.2 | 0.2 | 0.2 |
| 2013 | 1.326 | 0.962 | 0.233 | 0.2 | 0.2 | 0.2 |
| 2014* | 1.326 | 0.962 | 0.233 | 0.2 | 0.2 | 0.2 |
| 2015* | 1.326 | 0.962 | 0.233 | 0.2 | 0.2 | 0.2 |

*A new key run was performed in 2014 with data up to 2013 (ICES-WGSAM 2014), so 2014-2015 M-values are assumed equal to 2013.

Table 14.6 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Survey tuning indices for IBTS Q1 and Q3 (NS-IBTS Delta-GAM indices). Data used in the assessment are highlighted in bold font.

| IBTS_Q1_gam |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19832016 | 2016 |  |  |  |  |  |  |
|  |  |  | 0.25 |  |  |  |  |
| 1 | 5 | 0 |  |  |  |  |  |
| 1 | 3741.78 | 14132.13 | 1724.21 | 941.86 | 384.66 | 377.18 | 1983 |
| 1 | 11153.10 | 5497.36 | 2343.07 | 449.14 | 437.52 | 181.72 | 1984 |
| 1 | 544.18 | 14312.28 | 1967.92 | 741.63 | 240.04 | 271.17 | 1985 |
| 1 | 11081.92 | 2284.09 | 3338.45 | 883.19 | 449.11 | 250.91 | 1986 |
| 1 | 4343.54 | 13889.45 | 703.16 | 740.47 | 219.12 | 195.82 | 1987 |
| 1 | 2537.07 | 3376.75 | 3512.02 | 184.41 | 358.84 | 213.01 | 1988 |
| 1 | 8362.17 | 3247.15 | 2622.34 | 1092.07 | 154.26 | 242.90 | 1989 |
| 1 | 1770.07 | 7058.47 | 1148.25 | 427.25 | 487.25 | 79.49 | 1990 |
| 1 | 1537.86 | 2119.94 | 1874.37 | 492.86 | 271.26 | 264.44 | 1991 |
| 1 | 8175.95 | 2843.91 | 724.79 | 490.88 | 164.19 | 67.02 | 1992 |
| 1 | 2820.51 | 7559.90 | 931.68 | 324.35 | 248.62 | 76.59 | 1993 |
| 1 | 6187.10 | 2041.72 | 1528.11 | 465.48 | 236.70 | 125.76 | 1994 |
| 1 | 6769.18 | 8786.29 | 1633.26 | 539.54 | 158.52 | 71.91 | 1995 |
| 1 | 1639.53 | 4144.51 | 2101.74 | 325.73 | 234.19 | 81.02 | 1996 |
| 1 | 13784.83 | 2498.59 | 1156.42 | 506.59 | 169.80 | 115.96 | 1997 |
| 1 | 1578.96 | 9421.22 | 904.93 | 449.40 | 223.89 | 105.38 | 1998 |
| 1 | 1305.47 | 859.71 | 4033.23 | 362.04 | 212.58 | 87.47 | 1999 |
| 1 | 3481.66 | 2042.73 | 418.88 | 958.79 | 194.87 | 126.99 | 2000 |
| 1 | 920.90 | 3624.01 | 812.00 | 162.67 | 162.18 | 50.59 | 2001 |
| 1 | 2474.57 | 1411.18 | 1497.13 | 248.10 | 53.35 | 64.21 | 2002 |
| 1 | 364.51 | 2066.03 | 555.69 | 428.01 | 134.33 | 34.97 | 2003 |
| 1 | 2486.79 | 1139.14 | 1155.09 | 215.67 | 183.72 | 73.08 | 2004 |
| 1 | 943.49 | 1458.11 | 476.54 | 403.32 | 69.40 | 110.93 | 2005 |
| 1 | 3493.36 | 857.28 | 747.85 | 155.34 | 88.82 | 58.63 | 2006 |
| 1 | 1202.53 | 2644.71 | 701.81 | 232.04 | 87.09 | 67.69 | 2007 |
| 1 | 1910.62 | 1051.92 | 1194.02 | 299.30 | 208.04 | 51.91 | 2008 |
| 1 | 961.17 | 1775.10 | 873.44 | 398.22 | 130.54 | 75.95 | 2009 |
| 1 | 2508.88 | 1583.70 | 1191.47 | 330.12 | 208.62 | 85.72 | 2010 |
| 1 | 662.54 | 2898.17 | 628.04 | 359.33 | 237.31 | 140.38 | 2011 |
| 1 | 1343.47 | 1542.92 | 1963.47 | 442.23 | 250.25 | 86.58 | 2012 |
| 1 | 1404.72 | 1367.10 | 804.77 | 591.39 | 401.67 | 101.03 | 2013 |
| 1 | 2357.56 | 1723.23 | 790.62 | 302.22 | 370.83 | 115.97 | 2014 |
| 1 | 1510.25 | 3623.29 | 1278.42 | 465.31 | 205.98 | 162.89 | 2015 |
| 1 | 878.73 | 1107.19 | 2015.79 | 665.88 | 369.41 | 137.22 | 2016 |
| IBTS_Q3_gam |  |  |  |  |  |  |  |
| 1992 | 2015 |  |  |  |  |  |  |
| 1 | 1 | 0.50 | 0.75 |  |  |  |  |
| 1 | 4 |  |  |  |  |  |  |
| 1 | 16659.55 | 1668.03 | 379.05 | 328.82 | 109.27 | 39.70 | 1992 |
| 1 | 4734.55 | 4283.37 | 441.09 | 134.85 | 79.57 | 11.90 | 1993 |
| 1 | 16986.96 | 2205.09 | 917.32 | 154.23 | 44.60 | 34.19 | 1994 |
| 1 | 9089.09 | 6754.62 | 656.67 | 291.25 | 45.05 | 20.24 | 1995 |
| 1 | 4982.21 | 2881.26 | 972.41 | 208.60 | 111.74 | 13.60 | 1996 |
| 1 | 29172.47 | 1940.78 | 710.92 | 251.72 | 55.20 | 39.71 | 1997 |
| 1 | 845.78 | 8731.51 | 705.17 | 189.67 | 114.94 | 38.66 | 1998 |
| 1 | 3261.62 | 461.16 | 2348.48 | 152.85 | 40.78 | 17.98 | 1999 |
| 1 | 6320.26 | 915.81 | 111.73 | 334.77 | 35.15 | 32.31 | 2000 |
| 1 | 1380.10 | 2145.59 | 362.00 | 74.17 | 56.05 | 35.73 | 2001 |
| 1 | 3817.10 | 830.28 | 742.27 | 204.51 | 54.31 | 25.45 | 2002 |
| 1 | 980.98 | 1205.45 | 263.92 | 176.23 | 84.68 | 57.54 | 2003 |
| 1 | 3148.20 | 753.56 | 468.06 | 93.26 | 68.88 | 25.32 | 2004 |
| 1 | 1075.17 | 716.81 | 281.62 | 115.61 | 25.03 | 43.83 | 2005 |
| 1 | 5461.72 | 682.03 | 581.09 | 115.39 | 27.80 | 17.89 | 2006 |
| 1 | 1835.00 | 2275.05 | 425.92 | 167.70 | 93.09 | 43.57 | 2007 |
| 1 | 2455.73 | 1127.97 | 1084.04 | 219.96 | 114.25 | 30.28 | 2008 |
| 1 | 1917.58 | 922.39 | 291.62 | 234.33 | 50.64 | 25.19 | 2009 |
| 1 | 4570.25 | 1661.53 | 540.67 | 179.22 | 105.49 | 21.96 | 2010 |
| 1 | 1205.34 | 2767.19 | 864.61 | 358.71 | 98.39 | 95.45 | 2011 |
| 1 | 2064.47 | 974.01 | 1239.32 | 364.15 | 100.85 | 18.89 | 2012 |
| 1 | 3098.67 | 1073.44 | 501.20 | 527.55 | 141.27 | 63.70 | 2013 |
| 1 | 3391.89 | 1481.11 | 597.75 | 291.73 | 201.08 | 95.81 | 2014 |
| 1 | 1834.02 | 3110.39 | 1085.68 | 466.09 | 143.51 | 141.77 | 2015 |

Table 14.7a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. SAM final run model specification (model.cfg file).
\# Min Age (should not be modified unless data is modified accordingly)
1
\# Max Age (should not be modified unless data is modified accordingly)
6
\# Max Age considered a plus group ( $0=$ No, $1=\mathrm{Yes}$ )
1
\# The following matrix describes the coupling
\# of fishing mortality
\# Rows represent fleets.
\# Columns represent ages.

| 1 | 2 | 3 | 4 | 5 | 6 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 |

\# Use correlated random walks for the fishing mortalities
\# ( 0 = independent, 1 = correlation estimated $)$
2
\# Coupling of catchability PARAMETERS

| 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 2 | 3 | 4 | 5 | 0 |
| 6 | 7 | 8 | 9 | 0 | 0 |
| $\#$ |  |  |  |  |  |
| \# Coupling of power law model EXPONENTS (if used) |  |  |  |  |  |


| 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 |
| \# Coupling of fishing mortality |  |  |  |  |  |


| 1 | 2 | 2 | 2 | 2 | 2 |
| :--- | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 |
| \# Coupling of log N | RW VARIANCES |  |  |  |  |
| 1 | 2 | 2 | 2 | 2 | 2 |
| \# Coupling of OBSERVATION VARIANCES |  |  |  |  |  |
| 1 | 2 | 3 | 3 | 3 | 3 |
| 4 | 5 | 5 | 5 | 5 | 0 |
| 6 | 7 | 7 | 7 | 0 | 0 |

```
# Stock recruitment model code ( 0=RW, 1=Ricker, 2=BH, ... more in time)
0
# Years in which catch data are to be scaled by an estimated parameter
# first the number of years
1 3
# Then the actual years
1 9 9 3 1 9 9 4 1 9 9 5 1 9 9 6 1 9 9 7 1 9 9 8 1 9 9 9 2 0 0 0 2 0 0 1 2 0 0 2 2 0 0 3 2 0 0 4 2 0 0 5 ~
# Them the model config lines years cols ages
111111
222222
333333
444444
555555
666666
77777
888888
9 9 9 9 9 9
101010101010
111111111111
121212121212
131313131313
# Define Fbar range
2 4
```

Table 14.7b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. SAM final run model fitting diagnostics, parameter estimates and correlation matrix (.par and .cor files)
\# Number of parameters $=34$ Objective function value $=144.135$ Maximum gradient component $=$ 0.00930897

```
The logarithm of the determinant of the hessian = 164.414
    1 logqQ1_1
    2 logqQ1_2
    3 logq Q1-3
    4 logq Q1_4
    5 logqQ1-5
    6 logq Q3_1
    7 logq Q3_2
    8 logqQ3_3
    9 9 logq Q3_4
    11 logSd LogF2+
    12 logSd LogN1
    13 logSd LogN2+
    4 logSd LogC1
    15 logSd LogC2
    1 6 ~ l o g S d ~ L o g C 3 +
    17 logSd LogQ1-
    18 logSd LogQ1_2
    19 logSd LogQ3_
    O logSd LogQ3
    21 rho
    22 log Cmult 9
    2 log Cmult 94
    24 log Cmult 95
    25 log Cmult 96
    26 log Cmult 97
    27 log Cmult 98
    log Cmult 00
    30 log Cmult 01
    31 log Cmult 02
    32 log Cmult 03
    33 log Cmult 04
    34 log Cmult 05
```



```
    %
-1.4242
-0.77307
    1.4928
-1.0091
0.04718}0.054397 -0, 0, 0.01 0.03 0.07 rrrren\mp@code{0.02
0.1587 [0.09858 
0.012358
-0.15563 [10.09646
-0.34224 0.096725
```



```
0.25665
-0.13797 0.096256 -0.1 -0.2
```




Table 14.8 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and VIId. SAM final run estimated fishing mortality at age.

Fishing mortality ( F ) at age

| Year\Age | 1 | 2 | 3 | 4 | 5 | 6+ | Fbar 2-4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1963 | 0.079 | 0.443 | 0.511 | 0.465 | 0.461 | 0.473 | 0.473 |
| 1964 | 0.089 | 0.478 | 0.562 | 0.507 | 0.500 | 0.509 | 0.516 |
| 1965 | 0.104 | 0.528 | 0.623 | 0.549 | 0.531 | 0.533 | 0.567 |
| 1966 | 0.109 | 0.541 | 0.631 | 0.544 | 0.528 | 0.533 | 0.572 |
| 1967 | 0.119 | 0.575 | 0.673 | 0.580 | 0.571 | 0.574 | 0.609 |
| 1968 | 0.131 | 0.612 | 0.711 | 0.615 | 0.602 | 0.596 | 0.646 |
| 1969 | 0.123 | 0.586 | 0.671 | 0.581 | 0.576 | 0.572 | 0.613 |
| 1970 | 0.143 | 0.638 | 0.711 | 0.601 | 0.586 | 0.570 | 0.650 |
| 1971 | 0.182 | 0.737 | 0.800 | 0.668 | 0.644 | 0.617 | 0.735 |
| 1972 | 0.215 | 0.812 | 0.860 | 0.716 | 0.689 | 0.660 | 0.796 |
| 1973 | 0.222 | 0.815 | 0.836 | 0.692 | 0.661 | 0.625 | 0.781 |
| 1974 | 0.219 | 0.793 | 0.789 | 0.651 | 0.630 | 0.599 | 0.744 |
| 1975 | 0.251 | 0.861 | 0.848 | 0.696 | 0.673 | 0.631 | 0.802 |
| 1976 | 0.288 | 0.934 | 0.908 | 0.726 | 0.701 | 0.650 | 0.856 |
| 1977 | 0.273 | 0.896 | 0.866 | 0.682 | 0.675 | 0.631 | 0.815 |
| 1978 | 0.307 | 0.972 | 0.969 | 0.765 | 0.756 | 0.693 | 0.902 |
| 1979 | 0.283 | 0.908 | 0.917 | 0.713 | 0.689 | 0.628 | 0.846 |
| 1980 | 0.313 | 0.974 | 1.005 | 0.783 | 0.737 | 0.667 | 0.921 |
| 1981 | 0.312 | 0.979 | 1.029 | 0.803 | 0.738 | 0.668 | 0.937 |
| 1982 | 0.350 | 1.071 | 1.159 | 0.917 | 0.833 | 0.747 | 1.049 |
| 1983 | 0.340 | 1.057 | 1.149 | 0.920 | 0.826 | 0.737 | 1.042 |
| 1984 | 0.306 | 0.985 | 1.066 | 0.871 | 0.783 | 0.700 | 0.974 |
| 1985 | 0.285 | 0.945 | 1.022 | 0.852 | 0.762 | 0.680 | 0.940 |
| 1986 | 0.297 | 0.978 | 1.078 | 0.923 | 0.820 | 0.726 | 0.993 |
| 1987 | 0.280 | 0.952 | 1.058 | 0.918 | 0.812 | 0.718 | 0.976 |
| 1988 | 0.283 | 0.964 | 1.085 | 0.941 | 0.823 | 0.719 | 0.997 |
| 1989 | 0.290 | 0.979 | 1.101 | 0.962 | 0.844 | 0.734 | 1.014 |
| 1990 | 0.262 | 0.918 | 1.018 | 0.884 | 0.768 | 0.664 | 0.940 |
| 1991 | 0.251 | 0.896 | 1.007 | 0.890 | 0.785 | 0.673 | 0.931 |
| 1992 | 0.241 | 0.877 | 0.999 | 0.888 | 0.778 | 0.652 | 0.921 |
| 1993 | 0.240 | 0.883 | 1.027 | 0.907 | 0.790 | 0.649 | 0.939 |
| 1994 | 0.241 | 0.890 | 1.065 | 0.931 | 0.808 | 0.651 | 0.962 |
| 1995 | 0.249 | 0.923 | 1.130 | 0.973 | 0.842 | 0.662 | 1.009 |
| 1996 | 0.232 | 0.895 | 1.136 | 0.994 | 0.889 | 0.693 | 1.008 |
| 1997 | 0.212 | 0.847 | 1.113 | 0.996 | 0.902 | 0.690 | 0.985 |
| 1998 | 0.209 | 0.843 | 1.139 | 1.030 | 0.933 | 0.696 | 1.004 |
| 1999 | 0.215 | 0.863 | 1.210 | 1.113 | 1.016 | 0.741 | 1.062 |
| 2000 | 0.211 | 0.855 | 1.217 | 1.136 | 1.032 | 0.730 | 1.069 |
| 2001 | 0.190 | 0.798 | 1.133 | 1.068 | 0.963 | 0.665 | 1.000 |
| 2002 | 0.176 | 0.751 | 1.073 | 1.013 | 0.909 | 0.616 | 0.946 |
| 2003 | 0.171 | 0.735 | 1.056 | 0.977 | 0.862 | 0.568 | 0.923 |
| 2004 | 0.166 | 0.716 | 1.030 | 0.919 | 0.812 | 0.526 | 0.888 |
| 2005 | 0.153 | 0.674 | 0.963 | 0.839 | 0.758 | 0.486 | 0.825 |
| 2006 | 0.134 | 0.611 | 0.858 | 0.728 | 0.673 | 0.428 | 0.732 |
| 2007 | 0.119 | 0.561 | 0.798 | 0.672 | 0.619 | 0.384 | 0.677 |
| 2008 | 0.108 | 0.525 | 0.761 | 0.637 | 0.597 | 0.369 | 0.641 |
| 2009 | 0.102 | 0.505 | 0.746 | 0.628 | 0.587 | 0.352 | 0.626 |
| 2010 | 0.083 | 0.437 | 0.647 | 0.544 | 0.505 | 0.298 | 0.543 |
| 2011 | 0.062 | 0.356 | 0.527 | 0.446 | 0.416 | 0.246 | 0.443 |
| 2012 | 0.054 | 0.324 | 0.481 | 0.406 | 0.371 | 0.216 | 0.404 |
| 2013 | 0.052 | 0.314 | 0.472 | 0.394 | 0.352 | 0.199 | 0.393 |
| 2014 | 0.052 | 0.314 | 0.479 | 0.394 | 0.344 | 0.192 | 0.396 |
| 2015 | 0.050 | 0.305 | 0.466 | 0.383 | 0.339 | 0.19 | 0.385 |

Table 14.9 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. SAM final run estimated population numbers at age. [Note, the recruitment value in the final year relies on a single data point only, and is therefore considered preliminary only, and is ignored for projections.]

| Year\Age | 1 | 2 | 3 | 4 | 5 | $6+$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1963 | 487478 | 184241 | 20339 | 10330 | 8251 | 5229 | 715868 |
| 1964 | 802109 | 133653 | 51741 | 11535 | 5261 | 7059 | 1011358 |
| 1965 | 1042362 | 229120 | 40336 | 22948 | 6212 | 5602 | 1346580 |
| 1966 | 1271872 | 275130 | 67711 | 16487 | 9513 | 6184 | 1646897 |
| 1967 | 1074107 | 338067 | 72693 | 28739 | 8039 | 8202 | 1529847 |
| 1968 | 544161 | 289816 | 89859 | 28739 | 14327 | 7123 | 974025 |
| 1969 | 480220 | 141210 | 72620 | 34166 | 11439 | 10048 | 749703 |
| 1970 | 1561254 | 130875 | 38754 | 32370 | 15432 | 8536 | 1787221 |
| 1971 | 2034987 | 414571 | 33591 | 14955 | 16036 | 10721 | 2524861 |
| 1972 | 509406 | 518658 | 90490 | 12157 | 5981 | 12826 | 1149518 |
| 1973 | 740440 | 119372 | 105873 | 29912 | 4964 | 7216 | 1007777 |
| 1974 | 726505 | 177194 | 23861 | 36680 | 11073 | 5634 | 980947 |
| 1975 | 1236753 | 169397 | 36901 | 9666 | 16455 | 7232 | 1476404 |
| 1976 | 846614 | 285501 | 33996 | 13639 | 3812 | 9945 | 1193507 |
| 1977 | 2092772 | 171442 | 51226 | 10937 | 5137 | 6186 | 2337700 |
| 1978 | 1333077 | 443743 | 30792 | 18834 | 5598 | 4652 | 1836696 |
| 1979 | 1636385 | 271577 | 80258 | 9005 | 7374 | 3572 | 2008171 |
| 1980 | 2623448 | 313013 | 58806 | 24711 | 3982 | 4721 | 3028681 |
| 1981 | 1056001 | 494350 | 58747 | 17678 | 8882 | 3660 | 1639318 |
| 1982 | 1727179 | 191760 | 92042 | 17024 | 6599 | 5632 | 2040236 |
| 1983 | 946949 | 320937 | 33090 | 21245 | 5434 | 4533 | 1332188 |
| 1984 | 1709993 | 177371 | 51896 | 8047 | 6790 | 3887 | 1957984 |
| 1985 | 415817 | 332701 | 32958 | 14820 | 2806 | 4265 | 803367 |
| 1986 | 1861699 | 84881 | 57815 | 10212 | 5661 | 3037 | 2023305 |
| 1987 | 709276 | 387317 | 16365 | 15466 | 3034 | 3409 | 1134867 |
| 1988 | 490411 | 151752 | 70263 | 5234 | 4986 | 2410 | 725056 |
| 1989 | 827364 | 107689 | 30394 | 17499 | 1832 | 2993 | 987771 |
| 1990 | 327420 | 180593 | 20296 | 7972 | 5228 | 1729 | 543238 |
| 1991 | 374370 | 75132 | 30669 | 6111 | 2741 | 3047 | 492070 |
| 1992 | 856834 | 90400 | 14979 | 8947 | 2073 | 2140 | 975373 |
| 1993 | 434521 | 201995 | 17910 | 5104 | 2918 | 1661 | 664109 |
| 1994 | 1018661 | 107259 | 36534 | 5594 | 1778 | 1797 | 1171623 |
| 1995 | 596002 | 248451 | 23318 | 10440 | 1823 | 1379 | 881413 |
| 1996 | 372876 | 141775 | 39577 | 5710 | 3344 | 1449 | 564731 |
| 1997 | 1160081 | 96568 | 25848 | 9548 | 1868 | 1723 | 1295636 |
| 1998 | 141210 | 300740 | 20613 | 7003 | 3011 | 1280 | 473857 |
| 1999 | 251450 | 37272 | 54285 | 5519 | 2072 | 1623 | 352221 |
| 2000 | 457257 | 67914 | 8617 | 10507 | 1578 | 1150 | 547023 |
| 2001 | 167042 | 127389 | 14258 | 2296 | 2420 | 810 | 314215 |
| 2002 | 246965 | 48388 | 25540 | 4067 | 626 | 1033 | 326619 |
| 2003 | 123254 | 68186 | 10749 | 7506 | 1153 | 589 | 211437 |
| 2004 | 201793 | 37459 | 14371 | 3257 | 2172 | 650 | 259702 |
| 2005 | 154508 | 53960 | 8150 | 3628 | 1066 | 1171 | 222483 |
| 2006 | 358255 | 45388 | 12417 | 2326 | 1170 | 1027 | 420583 |
| 2007 | 168552 | 100609 | 10064 | 4366 | 1043 | 918 | 285552 |
| 2008 | 196025 | 46490 | 23742 | 3432 | 1780 | 1123 | 272592 |
| 2009 | 193300 | 54122 | 11460 | 8139 | 1539 | 1326 | 269886 |
| 2010 | 296262 | 55160 | 13354 | 4308 | 3591 | 1296 | 373971 |
| 2011 | 148153 | 80902 | 13808 | 4803 | 2007 | 2619 | 252292 |
| 2012 | 203211 | 41689 | 21047 | 6704 | 2297 | 2377 | 277325 |
| 2013 | 263024 | 52892 | 11685 | 9939 | 3509 | 2362 | 343411 |
| 2014 | 391601 | 66836 | 15902 | 5772 | 5136 | 3012 | 488259 |
| 2015 | 169058 | 104925 | 20228 | 7482 | 3032 | 5274 | 309999 |
| 2016 | 145365 | 41357 | 30152 | 10075 | 4168 | 5339 | 236456 |

Table 14.10 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. SAM final run estimated total removals at age (including catches due to unaccounted mortality)

| Year\Age | 1 | 2 | 3 | 4 | 5 | $6+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1963 | 21689 | 47321 | 7397 | 3507 | 2785 | 1800 |
| 1964 | 39807 | 36523 | 20249 | 4191 | 1892 | 2573 |
| 1965 | 60409 | 67833 | 17038 | 8865 | 2341 | 2115 |
| 1966 | 76726 | 82975 | 28889 | 6330 | 3569 | 2337 |
| 1967 | 70841 | 107087 | 32461 | 11577 | 3199 | 3276 |
| 1968 | 39211 | 96221 | 41719 | 12080 | 5934 | 2928 |
| 1969 | 32725 | 45374 | 32377 | 13780 | 4584 | 4005 |
| 1970 | 122578 | 44860 | 18000 | 13377 | 6262 | 3394 |
| 1971 | 200426 | 158055 | 16906 | 6681 | 6971 | 4522 |
| 1972 | 58513 | 211758 | 47777 | 5700 | 2731 | 5676 |
| 1973 | 87571 | 48850 | 54852 | 13694 | 2200 | 3072 |
| 1974 | 84669 | 71168 | 11903 | 16073 | 4741 | 2325 |
| 1975 | 161765 | 72533 | 19313 | 4442 | 7390 | 3100 |
| 1976 | 124033 | 130027 | 18572 | 6458 | 1762 | 4356 |
| 1977 | 289179 | 76321 | 27152 | 4957 | 2311 | 2650 |
| 1978 | 202805 | 209190 | 17506 | 9245 | 2724 | 2131 |
| 1979 | 229349 | 122602 | 44135 | 4208 | 3364 | 1526 |
| 1980 | 400713 | 147798 | 34190 | 12315 | 1905 | 2107 |
| 1981 | 160171 | 233935 | 34641 | 8964 | 4255 | 1635 |
| 1982 | 289526 | 95894 | 58070 | 9404 | 3426 | 2719 |
| 1983 | 154848 | 158832 | 20787 | 11760 | 2807 | 2168 |
| 1984 | 256145 | 83952 | 31245 | 4302 | 3386 | 1795 |
| 1985 | 58924 | 153169 | 19341 | 7808 | 1374 | 1929 |
| 1986 | 275709 | 39971 | 34989 | 5662 | 2910 | 1438 |
| 1987 | 100881 | 179333 | 9796 | 8549 | 1550 | 1601 |
| 1988 | 71083 | 70785 | 42689 | 2937 | 2568 | 1133 |
| 1989 | 123624 | 50701 | 18596 | 9953 | 959 | 1428 |
| 1990 | 45143 | 81528 | 11869 | 4302 | 2573 | 769 |
| 1991 | 50071 | 33360 | 17815 | 3313 | 1368 | 1368 |
| 1992 | 111035 | 39525 | 8658 | 4841 | 1029 | 940 |
| 1993 | 56449 | 88610 | 10525 | 2800 | 1463 | 726 |
| 1994 | 133226 | 47259 | 21929 | 3119 | 905 | 788 |
| 1995 | 80443 | 112005 | 14466 | 5981 | 953 | 612 |
| 1996 | 47448 | 62442 | 24637 | 3313 | 1811 | 664 |
| 1997 | 136189 | 40864 | 15893 | 5545 | 1021 | 788 |
| 1998 | 16444 | 126475 | 12828 | 4151 | 1681 | 588 |
| 1999 | 30170 | 15878 | 34902 | 3420 | 1218 | 779 |
| 2000 | 54090 | 28598 | 5549 | 6585 | 936 | 547 |
| 2001 | 17978 | 50808 | 8822 | 1390 | 1378 | 360 |
| 2002 | 24629 | 18370 | 15307 | 2386 | 344 | 435 |
| 2003 | 11953 | 25276 | 6376 | 4311 | 612 | 234 |
| 2004 | 19026 | 13525 | 8389 | 1802 | 1109 | 243 |
| 2005 | 13408 | 18522 | 4568 | 1892 | 520 | 412 |
| 2006 | 27203 | 14393 | 6467 | 1103 | 525 | 326 |
| 2007 | 11364 | 29747 | 5000 | 1957 | 441 | 267 |
| 2008 | 12003 | 13020 | 11428 | 1481 | 733 | 316 |
| 2009 | 11140 | 14653 | 5445 | 3477 | 626 | 359 |
| 2010 | 13774 | 13201 | 5742 | 1652 | 1302 | 304 |
| 2011 | 5121 | 16233 | 5093 | 1577 | 623 | 521 |
| 2012 | 6079 | 7658 | 7229 | 2042 | 650 | 419 |
| 2013 | 7376 | 9375 | 3958 | 2955 | 949 | 388 |
| 2014 | 10986 | 11854 | 5446 | 1713 | 1364 | 478 |
| 2015 | 4556 | 18129 | 6779 | 2172 | 794 | 830 |

Table 14.11a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. SAM final run estimated stock and management metrics, together with the lower and upper bounds of the point-wise $\mathbf{9 5 \%}$ confidence intervals.

Estimated recruitment, total stock biomass (TBS), spawning stock biomass (SSB), total removals (including catches due to unaccounted mortality) and average fishing mortality for ages 2 to 4 (Fbar 2-4).

| Year | Recruits age 1 ('000) | Low High | $\begin{array}{r} \text { TSB } \\ \text { (tons) } \\ \hline \end{array}$ |  | High | $\begin{array}{r} \text { SSB } \\ \text { (tons) } \\ \hline \end{array}$ | Low | High | Total removals (tons) | Low | High | Fbar 2-4 | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1963 | 487478 | 35832766379 | 511959 | 438909 | 597167 | 151903 | 116620 | 197861 | 117830 | 104453 | 132921 | 0.473 | 0.41 | 0.546 |
| 1964 | 802109 | 5907101089163 | 661986 | 563847 | 777207 | 163081 | 128015 | 207752 | 144929 | 131184 | 160114 | 0.515 | 0.453 | 0.587 |
| 1965 | 1042362 | 7700731410931 | 834009 | 718024 | 968731 | 199187 | 161242 | 246062 | 199187 | 177659 | 223324 | 0.567 | 0.498 | 0.645 |
| 1966 | 1271872 | 9407381719562 | 996496 | 858764 | 1156316 | 221239 | 180297 | 27479 | 240867 | 215473 | 269254 | 0.572 | 0.505 | 0.648 |
| 1967 | 1074107 | 7936361453696 | 1051786 | 916308 | 1207294 | 249946 | 204164 | 305994 | 287506 | 256711 | 321996 | 0.609 | 0.541 | 0.687 |
| 1968 | 544161 | 401563737396 | 879404 | 784009 | 986407 | 260928 | 219902 | 309608 | 293608 | 266422 | 323568 | 0.646 | 0.572 | 0.729 |
| 1969 | 480220 | 352625653985 | 735275 | 650773 | 830750 | 257816 | 215292 | 308739 | 226840 | 209423 | 245706 | 0.613 | 0.545 | 0.689 |
| 1970 | 1561254 | 11517352116385 | 1191829 | 989256 | 1435885 | 270222 | 226747 | 322033 | 252206 | 221705 | 286903 | 0.65 | 0.581 | 0.727 |
| 1971 | 2034987 | 14948552770283 | 1330413 | 1125138 | 1573440 | 274032 | 230490 | 325799 | 349410 | 300289 | 406565 | 0.735 | 0.66 | 0.819 |
| 1972 | 509406 | 373697694398 | 921723 | 816093 | 1041025 | 242559 | 204115 | 288243 | 362580 | 377364 | 44237 | 0.796 | 0.714 | 0.887 |
| 1973 | 740440 | 5434321008869 | 738222 | 654242 | 832983 | 209190 | 18178 | 241534 | 259627 | 236256 | 285309 | 0.781 | 0.701 | 0.87 |
| 1974 | 726505 | 532203991744 | 710696 | 628630 | 803476 | 227521 | 197185 | 262526 | 235626 | 210477 | 263779 | 0.744 | 0.668 | 0.829 |
| 1975 | 1236753 | 8980491703201 | 806936 | 687318 | 947373 | 208147 | 178992 | 242051 | 245242 | 213679 | 281467 | 0.802 | 0.722 | 0.89 |
| 1976 | 846614 | $610455 \quad 174132$ | 636029 | 558694 | 724069 | 177549 | 150733 | 209135 | 245242 | 212833 | 282586 | 0.856 | 0.77 | 0.952 |
| 1977 | 2092772 | 15778722885417 | 987567 | 804712 | 121973 | 152512 | 129838 | 179147 | 259367 | 213325 | 315347 | 0.815 | 0.733 | 0.906 |
| 1978 | 1333077 | 9643021842880 | 1126921 | 939155 | 1352228 | 153430 | 135157 | 174173 | 355045 | 291999 | 431704 | 0.902 | 0.814 | 1 |
| 1979 | 1636385 | 11873092255315 | 1040280 | 885778 | 1222558 | 155438 | 138265 | 174743 | 340102 | 290766 | 397808 | 0.846 | 0.764 | 0.937 |
| 1980 | 2623448 | 18947173632458 | 1255444 | 1040773 | 1544394 | 171785 | 153907 | 19740 | 391601 | 324430 | 472680 | 0.921 | 0.834 | 1.016 |
| 1981 | 1056001 | 7643341458968 | 1037163 | 896985 | 1199249 | 186839 | 168927 | 206650 | 395933 | 337515 | 464462 | 0.937 | 0.851 | 1.032 |
| 1982 | 1727179 | 12664402355539 | 1132570 | 947803 | 1353356 | 181861 | 163890 | 201804 | 386930 | 327705 | 456860 | 1.049 | 0.954 | 1.154 |
| 1983 | 946949 | 7053011271388 | 885582 | 761846 | 1029414 | 153737 | 138111 | 171131 | 324811 | 276910 | 380999 | 1.042 | 0.949 | 1.144 |
| 1984 | 1709993 | 12762542291140 | 908000 | 761709 | 1082388 | 132191 | 18318 | 147690 | 278173 | 236185 | 327626 | 0.974 | 0.887 | 1.07 |
| 1985 | 415817 | 306386564333 | 586542 | 519056 | 662802 | 134054 | 120046 | 149698 | 241832 | 209515 | 279134 | 0.939 | 0.854 | 1.033 |
| 1986 | 1861699 | 13921222489668 | 817495 | 671484 | 995255 | 117948 | 106247 | 130938 | 227749 | 190592 | 272150 | 0.993 | 0.905 | 1.089 |
| 1987 | 709276 | 531862945870 | 748630 | 644155 | 870050 | 124368 | 11732 | 138432 | 257558 | 217457 | 305054 | 0.976 | 0.889 | 1.071 |
| 1988 | 490411 | 367376654653 | 550730 | 481387 | 630062 | 122394 | 111714 | 134095 | 206076 | 182711 | 232429 | 0.997 | 0.909 | 1.093 |
| 1989 | 827364 | 674421108657 | 555154 | 469661 | 656209 | 109754 | 99560 | 120993 | 179154 | 154255 | 208071 | 1.014 | 0.924 | . 113 |
| 1990 | 327420 | 245925435923 | 371759 | 327466 | 422042 | 99409 | 89727 | 110136 | 138275 | 120905 | 158140 | 0.94 | 0.853 | 1.035 |
| 1991 | 374370 | 282954495320 | 342491 | 299039 | 392256 | 95607 | 85764 | 106579 | 118302 | 105025 | 133259 | 0.931 | 0.847 | 1.025 |
| 1992 | 856834 | 6515961126719 | 534988 | 446930 | 640396 | 91400 | 8247 | 101362 | 140084 | 18184 | 166044 | 0.921 | 0.838 | 1.012 |
| 1993 | 434521 | 333421566278 | 414571 | 364768 | 471175 | 98815 | 89788 | 108748 | 148301 | 127906 | 17948 | 0.939 | 0.855 | 1.032 |
| 1994 | 1018661 | 774196134032 | 529136 | 447527 | 625626 | 101722 | 173 | 111056 | 153430 | 132162 | 178121 | 0.962 | 0.878 | 1.054 |
| 1995 | 596002 | 455413779993 | 564107 | 487180 | 653183 | 121297 | 111138 | 132385 | 190232 | 162909 | 222137 | 1.009 | 0.92 | 05 |
| 1996 | 372876 | 286009486 124 | 420837 | 371918 | 476190 | 116658 | 107364 | 126756 | 155593 | 137976 | 175460 | 1.008 | 0.921 | 4 |
| 1997 | 1160081 | 8737591540226 | 643064 | 525791 | 786494 | 101519 | 93269 | 110499 | 153430 | 12837 | 183458 | 0.985 | 0.901 | 1.078 |
| 1998 | 141210 | $107355 \quad 185739$ | 328076 | 288902 | 372562 | 102847 | 93052 | 113674 | 135673 | 116389 | 158151 | 1.004 | 0.919 | 1.097 |
| 1999 | 251450 | 193327327047 | 226840 | 203275 | 253136 | 85819 | 78573 | 93734 | 94845 | 86633 | 103836 | 1.062 | 0.972 | 1.61 |
| 2000 | 457257 | 351456594907 | 289526 | 246955 | 339435 | 68255 | 62000 | 75140 | 84965 | 73561 | 98138 | 1.069 | 0.978 | 1.17 |
| 2001 | 167042 | $128335 \quad 217423$ | 198590 | 176133 | 223911 | 63513 | 57588 | 70047 | 72186 | 63669 | 81843 | 1 | 0.911 | 1.098 |
| 2002 | 246965 | 190137320777 | 168890 | 148613 | 191932 | 56387 | 51162 | 62146 | 56444 | 51043 | 62416 | 0.946 | 0.86 | 1.04 |
| 2003 | 123254 | $94570 \quad 160637$ | 142344 | 127825 | 158512 | 56783 | 51588 | 62503 | 53316 | 47825 | 59439 | 0.923 | 0.834 | 1.022 |
| 2004 | 201793 | 155640261634 | 123995 | 108845 | 141254 | 46212 | 41454 | 51516 | 39419 | 35863 | 43328 | 0.888 | 0.801 | 0.985 |
| 2005 | 154508 | 17922202444 | 139107 | 121516 | 159244 | 47620 | 41866 | 54164 | 40055 | 35433 | 45280 | 0.825 | 0.742 | 0.918 |
| 2006 | 358255 | 276697463852 | 146679 | 123856 | 173707 | 43261 | 37523 | 49876 | 31761 | 28205 | 35766 | 0.732 | 0.652 | 0.822 |
| 2007 | 168552 | $130595 \quad 217540$ | 195048 | 172164 | 220974 | 72766 | 64232 | 82433 | 53104 | 46487 | 60662 | 0.677 | 0.6 | 0.763 |
| 2008 | 196025 | 151700253303 | 205664 | 180468 | 234378 | 81227 | 71691 | 92031 | 52313 | 47548 | 57555 | 0.641 | 0.564 | 0.728 |
| 2009 | 193300 | 149561249831 | 220356 | 193500 | 250939 | 90944 | 79475 | 104068 | 54830 | 49572 | 60647 | 0.627 | 0.548 | 0.716 |
| 2010 | 296262 | 228415384262 | 236097 | 203918 | 273355 | 93060 | 79738 | 108608 | 48728 | 44262 | 53644 | 0.542 | 0.469 | 0.627 |
| 2011 | 148153 | 14403191860 | 223910 | 194119 | 258273 | 105662 | 88341 | 126379 | 44802 | 40440 | 49634 | 0.443 | 0.379 | 0.518 |
| 2012 | 203211 | 157358262425 | 199586 | 172234 | 231281 | 106831 | 88281 | 129277 | 40336 | 37275 | 43648 | 0.404 | 0.344 | 0.473 |
| 2013 | 263024 | 203288 340313 | 259886 | 223095 | 302746 | 117477 | 97190 | 142000 | 41606 | 38211 | 45303 | 0.393 | 0.338 | 0.458 |
| 2014 | 391601 | $295327 \quad 519260$ | 329391 | 278089 | 390156 | 126880 | 105516 | 152570 | 45936 | 41634 | 50682 | 0.395 | 0.341 | 0.459 |
| 2015 | 169058 | 11872241857 | 288082 | 245194 | 338472 | 151146 | 125031 | 182714 | 52313 | 46831 | 58436 | 0.385 | 0.327 | 0.453 |
| 2016 |  |  |  |  |  | 161135 | 129713 | 200170 |  |  |  |  |  |  |

Table 14.11b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. SAM final run estimated landings, discards, catch (=landings + discards) and total removals in tonnes. Landings and discards are derived by applying the landing fraction from landings and discards data to the SAM estimate of catch (after removing unaccounted mortality), while total removals are the SAM estimate of catch, including a catch multiplier incorporated from 1993 to 2005 only.

|  |  |  |  | Catch | Total |
| ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Landings | Discards | Catch multiplier | Removals |  |
| 1963 | 106938 | 10849 | 117830 | 117830 |  |
| 1964 | 135131 | 9788 | 144929 |  | 144929 |
| 1965 | 182043 | 17057 | 199187 |  | 199187 |
| 1966 | 214701 | 26239 | 240867 |  | 240867 |
| 1967 | 260667 | 26662 | 287506 |  | 287506 |
| 1968 | 276509 | 17127 | 293608 |  | 293608 |
| 1969 | 217075 | 9664 | 226840 |  | 226840 |
| 1970 | 232350 | 19897 | 252206 |  | 252206 |
| 1971 | 291560 | 57873 | 349410 |  | 349410 |
| 1972 | 328076 | 34372 | 362580 |  | 362580 |
| 1973 | 234685 | 24959 | 259627 |  | 259627 |
| 1974 | 209609 | 26160 | 235626 |  | 235626 |
| 1975 | 209190 | 36243 | 245242 |  | 245242 |
| 1976 | 201390 | 43871 | 245242 |  | 245242 |
| 1977 | 181680 | 77964 | 259367 |  | 259367 |
| 1978 | 306202 | 48728 | 355045 |  | 355045 |
| 1979 | 278173 | 62131 | 340102 |  | 340102 |
| 1980 | 290977 | 100912 | 391601 |  | 391601 |
| 1981 | 342148 | 53960 | 395933 |  | 395933 |
| 1982 | 323191 | 63577 | 386930 |  | 386930 |
| 1983 | 287794 | 37235 | 324811 |  | 324811 |
| 1984 | 209819 | 68050 | 278173 |  | 278173 |
| 1985 | 213844 | 28029 | 241832 |  | 241832 |
| 1986 | 168890 | 59042 | 227749 |  | 227749 |
| 1987 | 225032 | 32565 | 257558 |  | 257558 |
| 1988 | 191377 | 14707 | 206076 |  | 206076 |
| 1989 | 138968 | 40296 | 179154 |  | 179154 |
| 1990 | 115151 | 23086 | 138275 |  | 138275 |
| 1991 | 102437 | 15755 | 118302 |  | 118302 |
| 1992 | 108554 | 31414 | 140084 |  | 140084 |
| 1993 | 130115 | 28543 | 158606 | 0.94 | 148301 |
| 1994 | 106116 | 41910 | 148048 | 1.04 | 153430 |
| 1995 | 130522 | 31930 | 162316 | 1.17 | 190232 |
| 1996 | 132275 | 21451 | 153682 | 1.01 | 155593 |
| 1997 | 133070 | 46149 | 179267 | 0.86 | 153430 |
| 1998 | 147449 | 43575 | 191040 | 0.71 | 135673 |
| 1999 | 96722 | 13843 | 110592 | 0.86 | 94845 |
| 2000 | 73373 | 16493 | 89798 | 0.95 | 84965 |
| 2001 | 44416 | 11411 | 55846 | 1.29 | 72186 |
| 2002 | 53422 | 11395 | 64794 | 0.87 | 56444 |
| 2003 | 31131 | 4750 | 35881 | 1.49 | 53316 |
| 2004 | 27269 | 7503 | 34770 | 1.13 | 39419 |
| 2005 | 29902 | 11366 | 41262 | 0.97 | 40055 |
| 2006 | 22629 | 9121 | 31761 |  | 31761 |
| 2007 | 24005 | 29144 | 53104 | 53104 |  |
| 2008 | 27038 | 25261 | 52313 |  | 52313 |
| 2009 | 33223 | 21610 | 54830 |  | 54830 |
| 2010 | 36207 | 12545 | 48728 |  | 48728 |
| 2011 | 34372 | 10443 | 44802 |  | 44802 |
| 2012 | 32728 | 7632 | 40336 | 40336 |  |
| 2013 | 30822 | 10808 | 41606 | 41606 |  |
| 2014 | 34822 | 11121 | 45936 | 45936 |  |
| 2015 | 38638 | 13654 | 52313 | 52313 |  |
|  |  |  |  |  |  |

Table 14.11c Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. SAM final run estimated catch multipliers, together with the lower and upper bounds of the point-wise $95 \%$ confidence intervals.
Year Catch multiplier

| year | Catch <br> multiplier | Low | High |
| ---: | ---: | ---: | ---: |
| 1993 | 0.94 | 0.78 | 1.12 |
| 1994 | 1.04 | 0.85 | 1.26 |
| 1995 | 1.17 | 0.96 | 1.43 |
| 1996 | 1.01 | 0.83 | 1.23 |
| 1997 | 0.86 | 0.71 | 1.04 |
| 1998 | 0.71 | 0.59 | 0.86 |
| 1999 | 0.86 | 0.70 | 1.05 |
| 2000 | 0.95 | 0.78 | 1.15 |
| 2001 | 1.29 | 1.06 | 1.57 |
| 2002 | 0.87 | 0.72 | 1.06 |
| 2003 | 1.49 | 1.22 | 1.81 |
| 2004 | 1.13 | 0.93 | 1.38 |
| 2005 | 0.97 | 0.81 | 1.16 |

Table 14.12a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Catch options based on the final SAM assessment run with maturities smoothed to 2015, where SSB in the intermediate year is 161135 tonnes. Units are '000t (SSB, landings, discards, unaccounted) or millions (recruitment).

| $\begin{aligned} & \text { Intermediate year F } \\ & \text { Recruitment resamp } \\ & \text { SSB(2017) = } \\ & \text { HC landings (2016) = } \\ & \text { Discards (2016) = } \\ & \hline \end{aligned}$ | assumpti led from | f: F(2016) 1998-2015 | $=F(2015)=$ | $\begin{aligned} & 0.39 \\ & 196 \\ & 174300 \\ & 44837 \\ & 11465 \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Rationale | $\begin{aligned} & \hline \text { Catch } \\ & \text { (2017) } \\ & \hline \end{aligned}$ | Landings (2017) | $\begin{gathered} \hline \text { Discards } \\ (2017) \\ \hline \end{gathered}$ | Basis | $\begin{aligned} & \hline \text { Ftotal } \\ & (2017) \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \text { Fland } \\ & \text { (2017) } \end{aligned}$ | $\begin{aligned} & \hline \text { Fdisc } \\ & \text { (2017) S } \end{aligned}$ | SSB (2018) | $\begin{gathered} \hline \text { SSB 5\% } \\ (2018) \\ \hline \end{gathered}$ | $\begin{gathered} \text { \%SSB } \\ \text { change } \end{gathered}$ | $\begin{array}{r} \hline \% \text { TAC } \\ \text { change } \end{array}$ | $\begin{aligned} & \hline \text { Ftotal } \\ & (2018) \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \text { Ftotal } \\ & (2019) \\ & \hline \end{aligned}$ | $\begin{gathered} \hline \text { Catch } \\ (2018) \\ \hline \end{gathered}$ | $\begin{aligned} & \hline \text { Catch } \\ & (2019) \\ & \hline \end{aligned}$ | Landings (2018) | $\begin{array}{r} \hline \text { Landings } \\ (2019) \end{array}$ | $\begin{array}{r} \hline \text { Discards } \\ (2018) \\ \hline \end{array}$ | $\begin{array}{r} \hline \text { Discards } \\ (2019) \\ \hline \end{array}$ | $\begin{array}{r} \text { SSB } \\ (2019) \\ \hline \end{array}$ | $\begin{array}{r} \text { SSB } \\ (2020) \\ \hline \end{array}$ | \%change SSB 19:17 | $\begin{aligned} & \hline \text { \%change } \\ & \text { SSB 20:17 } \end{aligned}$ |
| Management Plan | 55959 | 45612 | 10347 | EUMP | 0.40 | 0.28 | 0.12 | 173495 | 134519 | - | 13 | 0.40 | 0.40 | 53176 | 51087 | 42900 | 39989 | 10276 | 11098 | 171376 | 168355 | -2 | -3 |
| Management Plan | 54046 | 44091 | 9955 | EU-Norway | 0.38 | 0.27 | 0.11 | 175637 | 136324 | 1 | 9 | 0.40 | 0.40 | 53806 | 51541 | 43466 | 40319 | 10340 | 11222 | 173013 | 169678 | -1 | -3 |
| MSY approach | 47431 | 38691 | 8740 | FMSY *SB2017/Btrigger | 0.33 | 0.23 | 0.10 | 182807 | 142585 | 5 | -4 | 0.33 | 0.33 | 47361 | 47155 | 38465 | 37311 | 8896 | 9844 | 188515 | 191706 | 8 | 10 |
| Zero Catch | 0 | 0 | 0 | $\mathrm{F}=0$ | 0.00 | 0.00 | 0.00 | 237118 | 188285 | 36 | -100 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 307740 | 372686 | 77 | 114 |
| MSY | 47431 | 38691 | 8740 | FMSY | 0.33 | 0.23 | 0.10 | 182807 | 142585 | 5 | -4 | 0.33 | 0.33 | 47361 | 47155 | 38465 | 37311 | 8896 | 9844 | 188515 | 191706 | 8 | 10 |
| Fpa 0.41 | 57140 | 46551 | 10589 | Flim/1.4 | 0.41 | 0.29 | 0.12 | 172171 | 133416 | -1 | 15 | 0.41 | 0.41 | 53895 | 51541 | 43421 | 40320 | 10474 | 11221 | 169103 | 165281 | -3 | -5 |
| Fpa 0.47 | 63993 | 52118 | 11875 | $\mathrm{F}_{\text {lim }} \times \exp (-\sigma \times 1.645)$ | 0.47 | 0.33 | 0.14 | 164628 | 127030 | -6 | 29 | 0.47 | 0.47 | 58068 | 54061 | 46306 | 41819 | 11762 | 12242 | 155630 | 148788 | -11 | -15 |
| Flim | 75810 | 61629 | 14181 | Flim | 0.58 | 0.41 | 0.17 | 151846 | 116043 | -13 | 52 | 0.58 | 0.58 | 63649 | 56608 | 50331 | 43092 | 13318 | 13516 | 134929 | 123262 | -23 | -29 |
| SSB(2018) $=$ Blim | 107401 | 87011 | 20390 | SSB(2018)=Blim | 0.94 | 0.66 | 0.28 | 118000 | 88722 | -32 | 115 | 0.94 | 0.94 | 72088 | 56978 | 54797 | 40803 | 17291 | 16175 | 88170 | 72950 | -49 | -58 |
| SSB(2018)=Bpa | 63653 | 51839 | 11814 | SSB(2018)=Bpa | 0.47 | 0.33 | 0.14 | 165000 | 127386 | -5 | 28 | 0.47 | 0.47 | 57877 | 53933 | 46163 | 41770 | 11714 | 12163 | 156282 | 149569 | -10 | -14 |
| SSB(2018)=Btrigger | 63653 | 51839 | 11814 | SSB(2018)=Brrigger | 0.47 | 0.33 | 0.14 | 165000 | 127386 | -5 | 28 | 0.47 | 0.47 | 57877 | 53933 | 46163 | 41770 | 11714 | 12163 | 156282 | 149569 | -10 | -14 |
| TAC constraint | 39518 | 32335 | 7183 | TAC2016-20\% | 0.27 | 0.19 | 0.08 | 191608 | 144116 | 10 | -20 | 0.26 | 0.25 | 39895 | 40043 | 32335 | 32335 | 7560 | 7708 | 207709 | 218101 | 19 | 25 |
| TAC constraint | 41995 | 34356 | 7639 | TAC2016-15\% | 0.29 | 0.20 | 0.09 | 188858 | 141560 | 8 | -15 | 0.28 | 0.28 | 42456 | 42690 | 34356 | 34356 | 8100 | 8334 | 201598 | 208834 | 16 | 20 |
| TAC constraint | 44478 | 36377 | 8101 | TAC2016-10\% | 0.31 | 0.22 | 0.09 | 186052 | 139029 | 7 | -10 | 0.3 | 0.31 | 45026 | 45374 | 36377 | 36377 | 8649 | 8997 | 195584 | 198896 | 12 | 14 |
| TAC constraint | 46962 | 38398 | 8564 | TAC2016-5\% | 0.33 | 0.23 | 0.10 | 183206 | 136365 | 5 | -5 | 0.33 | 0.34 | 47616 | 48076 | 38398 | 38398 | 9218 | 9678 | 189451 | 189490 | 9 | 9 |
| TAC constraint | 49454 | 40419 | 9035 | TAC2016 | 0.35 | 0.24 | 0.11 | 180305 | 133707 | 3 | 0 | 0.35 | 0.37 | 50214 | 50814 | 40419 | 40419 | 9795 | 10395 | 183390 | 180231 | 5 | 3 |
| TAC constraint | 51939 | 42440 | 9499 | TAC2016 + 5\% | 0.37 | 0.26 | 0.11 | 177556 | 131054 | 2 | 5 | 0.38 | 0.41 | 52811 | 53576 | 42440 | 42440 | 10371 | 11136 | 177101 | 170508 | 2 | -2 |
| TAC constraint | 54425 | 44461 | 9964 | TAC2016 + 10\% | 0.39 | 0.27 | 0.12 | 174736 | 128407 | 0 | 10 | 0.41 | 0.45 | 55444 | 56375 | 44461 | 44461 | 10983 | 11914 | 171028 | 160715 | -2 | -8 |
| TAC constraint | 56914 | 46482 | 10432 | TAC2016+15\% | 0.41 | 0.29 | 0.12 | 171902 | 125766 | -1 | 15 | 0.44 | 0.5 | 58084 | 59233 | 46482 | 46482 | 11602 | 12751 | 165080 | 151676 | -5 | -13 |
| TAC constraint | 59410 | 48503 | 10907 | TAC2016+20\% | 0.43 | 0.30 | 0.13 | 169100 | 123082 | -3 | 20 | 0.47 | 0.55 | 60721 | 62110 | 48503 | 48503 | 12218 | 13607 | 159094 | 142343 | -9 | -18 |
| Status quo | 54214 | 44226 | 9988 | Fsq | 0.39 | 0.27 | 0.12 | 175459 | 136166 | 1 | 9 | 0.39 | 0.39 | 52065 | 50407 | 42063 | 39461 | 10002 | 10946 | 174721 | 172896 | 0 | -1 |

Table 14.12b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Catch options based on the SAM sensitivity run with maturities smoothed to 2016, where SSB in the intermediate year is 146386 tonnes. Units are ' 000 t (SSB, landings, discards, unaccounted) or millions (recruitment).



Figure 14.1 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d: (a) stacked area plot of reported landings and estimated discards (in tons); (b) proportion of total numbers caught at age that are discarded; (c) proportion of total weight caught that is discarded; (d) and proportion of the total numbers caught that are discarded.


Figure 14.2a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d: Mean weight at age in the catch for ages 1-9.


Figure 14.2b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d: Annually varying maturity-at-age smoothed to 2015 (solid line) or 2016 (dashed line). Values for 1963-1972 are the former constant maturity values used for cod.


Figure 14.2c Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d: Smoothed, annually varying natural mortality from the 2014 key run (ICES-WGSAM 2014). Values for 1963-1972 are set equal to the 1973 value, while 2014 and 2015 are set equal to 2013.


Figure 14.3a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Distribution charts of cod ages 1-3+ caught in the IBTS Q1 survey $1997-2016$ in the North Sea.


Figure 14.3a contd. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Distribution charts of cod ages 1-3+ caught in the IBTS Q1 survey 1997-2016 in the North Sea


Figure 14.3a contd. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Distribution charts of cod ages 1-3+ caught in the IBTS Q1 survey 1997-2016 in the North Sea.


Figure 14.3a contd. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Distribution charts of cod ages 1-3+ caught in the IBTS Q1 survey 1997-2016 in the North Sea.


Figure 14.3b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Distribution charts of cod ages 1-3+ caught in the IBTS Q3 survey 1997-2015 in the North Sea.


Figure 14.3b contd. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Distribution charts of cod ages 1-3+ caught in the IBTS Q3 survey 1997-2015 in the North Sea.


Figure 14.3b contd. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Distribution charts of cod ages 1-3+ caught in the IBTS Q3 survey 1997-2015 in the North Sea.


Figure 14.3b contd. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Distribution charts of cod ages 1-3+ caught in the IBTS Q3 survey 1997-2015 in the North Sea.


Figure 14.3c Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Extension of cod standard area used for the NS-IBTS extended index. Crosses indicate suggested extensions to the survey (ICESWKROUND, 2009; ICES-WKCOD, 2011); green squares (light and dark) indicate where the IBTS group indicate data is available; yellow squares indicate where intermittent coverage does not allow inclusion and the IBTS WG considered should be omitted; light green squares indicate the recommended extension around Shetland (ICES-WKCOD, 2011).


Figure 14.3d Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Comparison of the Q1 and Q3 NS-IBTS extended indices to the corresponding NS-IBTS Delta GAM indices used in the assessment. The indices are mean-standardised with an offset for ease of presentation.


Figure 14.4a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Log mean standardised indices plotted by year (top left) and cohort (top right), log abundance curves (bottom left) and associated negative gradients for each cohort across the reference fishing mortality of age 2-4 (bottom right), for the IBTSQ1 groundfish survey (NS-IBTS Delta-GAM index).


Figure 14.4b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Log mean standardised indices plotted by year (top left) and cohort (top right), log abundance curves (bottom left) and associated negative gradients for each cohort across the reference fishing mortality of age 2-4 (bottom right), for the IBTSQ3 groundfish survey (NS-IBTS Delta-GAM index).


Figure 14.5a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Within-survey correlations for IBTSQ1 (NS-IBTS Delta-GAM index) for the period 1983-2016. Individual points are given by cohort (year-class), the solid line is a standard linear regression line, the broken line nearest to it a robust linear regression line, and "cor" denotes the correlation coefficient. The pair of broken lines on either side of the solid line indicate prediction intervals. The most recent data point appears in square brackets.


Figure 14.5b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Within-survey correlations for IBTSQ3 (NS-IBTS Delta-GAM index) for the period 1992-2015. Individual points are given by cohort (year-class), the solid line is a standard linear regression line, the broken line nearest to it a robust linear regression line, and "cor" denotes the correlation coefficient. The pair of broken lines on either side of the solid line indicate prediction intervals. The most recent data point appears in square brackets.


Age 5


Figure 14.5c Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Between-survey correlations for IBTSQ1 and Q3 surveys (NS-IBTS Delta-GAM indices) for the period 1992-2015. Individual points are given by cohort (year-class), the solid line is a standard linear regression line, and the broken line nearest to it a robust linear regression line. The pair of broken lines on either side of the solid line indicate prediction intervals. The most recent data appear in square brackets.


Figure 14.6 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. SURBAR summary plots for estimates of total mortality, spawning stock biomass, total biomass and recruitment for a combined SURBAR run with both surveys (IBTSQ1 and Q3 NS-IBTS Delta-GAM indices, ages 1-5). The smoothing parameter $l$ is set to 3 , and reference age at 3 . The shaded area represents $90 \%$ confidence bounds.


Figure 14.7 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Total catch-at-age matrix expressed as (a) numbers-at-age and (b) proportions-at-age, which have been standardised over time (for each age, this is achieved by subtracting the mean proportion-at-age over the time series, and dividing by the corresponding variance). Grey bubbles indicate proportions above the mean over the time series at each age.


## Ages 2 to 4



Figure 14.8 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Log-catch cohort curves (top panel) and the associated negative gradients for each cohort across the reference fishing mortality of age 2-4.


Figure 14.9a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Estimated SSB, F (2-4), recruitment (age 1) and the catch multiplier from the SAM assessment with maturities smoothed to 2015 (solid black lines=estimate and shaded area=corresponding point-wise $95 \%$ confidence intervals). The final SAM assessment for last year (2015) is plotted in light grey for the SSB, F and recruitment plots for comparison.


Figure 14.9b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Estimated SSB from the SAM assessment with maturities smoothed to 2016 (solid black lines=estimate and shaded area=corresponding point-wise $95 \%$ confidence intervals). The SAM assessment for last year (2015) is plotted in light grey.


Figure 14.10 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Normalized residuals for the SAM assessment, for total catch, IBTSQ1, IBTSQ3, and the recruitment and survival process error. Empty circles indicate a positive residual and filled circles negative residual.


Figure 14.11 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Retrospective estimates (10 years) from the SAM assessment. Estimated yearly SSB (top), average fishing mortality (middle) and recruitment age 1 (bottom), together with corresponding point-wise $95 \%$ confidence intervals. Mohn's $r$ given in each plot is calculated as: $=\frac{1}{n} \sum_{y=Y-n}^{Y-1}\left(1-X_{y, y} / X_{y, Y}\right)$, where the first subscript indicates the year $X$ (SSB, F or R) pertains to, and the second subscript the final full data year for the given assessment, with $Y$ indicating the most recent assessment.


Figure 14.12 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Anticlockwise from top left, point-wise estimates and $95 \%$ confidence intervals of spawning stock biomass (SSB), total stock biomass (TSB), recruitment (R(age 1)), the catch multiplier, catch and mean fishing mortality for ages 2-4 ( $\mathrm{F}(2-4$ ) ), from the SAM final run (catch multiplier estimated for 1993-2005 only). The heavy lines represent the point-wise estimate, and the light lines point-wise $95 \%$ confidence intervals. The open diamonds given in the catch plot represent model estimates of the total catch excluding unaccounted mortality, while the solid lines represent the total catch including unaccounted mortality for 1993-2005. The horizontal broken lines in the SSB plot indicate Blim=118 000t and Bpa=165 000t, and in the Fbar plot Flim=0.58, Fpa=0.41 and Fmsy=0.33. The horizontal broken line in the catch multiplier plot indicates a multiplier of 1. Catch, SSB and TSB are in tons, and R in thousands.


Figure 14.13 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. SAM estimates of fishing mortality. The top panel shows mean fishing mortality for ages 2-4 (shown in Figure 14.12), but split into landings and discards components by using ratios calculated from the landings and discards numbers at age from the reported catch data, while the bottom panel shows fishing mortality for each age.


Figure 14.14 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Comparison of final SAM assessment for 2016 with the final SAM assessment for 2015. Plots are as described in Figure 14.12.


Figure 14.15 Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Estimates of the number of 5-year-old and older cod in the population (solid line; thousands) and the percentage of 1 year olds by number that have survived to age 5 in the given year (hashed line).


Figure 14.16a Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Biomass indices by subregion (see Figure 14.16c), based on NS-IBTS Q1 and Q3 data. The biomass indices are derived by fitting a non-stationary Delta-GAM model (including ship effects) to numbers-at-age for the entire dataset and integrating the fitted abundance surface over each of the Subareas to obtain indices-at-age by area. These are then multiplied by smoothed weight-at-age estimates and summed to get the biomass indices.


Figure 14.16b Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Recruitment indices by subregion (see Figure 14.16c), based on NS-IBTS Q1 and Q3 data.


Figure 14.16c Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Subregions used to derive area-specific biomass indices based on NS-IBTS Q1 and Q3 data.

## 15 Pollack (Pol/achius pol/achius) in Subarea 4 and Division 3.a (North Sea and Skagerrak)

### 15.1 General Biology

The existing knowledge of pollack biology is summarised in the Stock Annex. According to this information it is benthopelagic, and is found down to 200 m . In Skagerrak, 0 -group pollack are regularly found in shallow areas close to the shore. Pollack are therefore protected from the fisheries in the early life stages. Pollack move gradually away from the coast into deeper waters as they grow.

Spawning takes place from January to May, depending on the area, and mostly at 100 $m$ depth. FAO reports maximum length at 130 cm and maximum weight at 18.1 kg . Female length-at-maturity is estimated at $>35 \mathrm{~cm}$, at 3-4years of age and growth after age 3 is about 7 cm per year (Heino et al. 2012). Pollack feeds mainly on fish, and incidentally on crustaceans and cephalopods.

### 15.2 Stock identity and possible assessment areas

WGNEW (ICES, 2012) proposed, based on a pragmatic approach, to distinguish three different stock units: the southern European Atlantic shelf (Bay of Biscay and Iberian Peninsula), the Celtic Seas, and the North Sea (including 7.d and 3.a). In the ICES advice, it was, however, decided to include 7.d Pollack in the Celtic Seas Ecoregion.

### 15.3 Management

For 4 and 3.a there are no formal TACs for pollack, but catches of pollack should be counted against the quota for some other species when caught in Norwegian waters south of $62^{\circ} \mathrm{N}$. There is a Minimum Landing Size of 30 cm in European Member States (Council Regulation (EU) 850/1998). No explicit objective has been defined, no precautionary reference points have been proposed, and there is no management plan. Analytical assessments leading to fisheries advice have never been carried out for pollack.

### 15.4 Fisheries data

Landings statistics for pollack are available from ICES, but are clearly incomplete in earlier years. From 1977 the data series appears to be reasonably consistent and adequate for allocating catches at least to ICES subareas. Considering that pollack is not subject to TAC regulations, a major incentive for mis- or underreporting is not present and landings figures are thus probably reflecting main trends in landings in the different areas.

Landings by country for the years 1977-2015 in Division 3.a (Skagerrak/Kattegat) and Subarea 4 (North Sea) are shown in Table 15.1. Figure 15.1 shows total landings in Subarea 4 and Division 3.a from 1977-2015. Two periods with high landings can be seen, and over the entire period total landings for both areas have declined. In Division 3.a landings have been low but stable since 2000, while in Subarea 4 landings have fluctuated over the same period and stabilised the last four years. Swedish fishers targeted pollack from the 1940s until mid-1980s when landings sometimes amounted to over 1000 tonnes. From the 1980s pollack started to decline severely and is today seldom caught in the Kattegat or along the Swedish Skagerrak coast.

Nowadays, no fishery is targeting pollack, and it is mainly, possibly exclusively, a bycatch in various commercial fisheries. Norwegian catches peak in the months of March
and April, and this may be associated with spawning aggregations. In 2015, 47\% of the total landings were caught with gillnet and $39 \%$ with otter trawls in Division 3.a. In Subarea $418 \%$ of the total landings were made with gillnets and $72 ? \%$ with otter trawls. The geographical distribution of Norwegian otter trawl catches resembles those of the saithe fisheries, but the catches of pollack are much lower. Discards are now considered by ICES to be known to take place and raised discards were estimated at 5.7 tonnes in total between area 3 and 4 in 2015 (see Table 15.2 for total catches and Table 15.3 for estimated discards). Discard numbers were raised for all nations. $72 \%$ of the discards were reported by bottom trawl fleets with UK-Scotland the country reporting the largest number of discards ( $30 \%$ of total).

Pollack is also frequently caught in recreational fisheries. Regularly collected data about these catches are not available to the working group. Norwegian recreational fishing data collected in 2009 suggests that catches of pollack south of $62^{\circ}$ north may range between 13 - 30 tonnes (Vølstad et al. 2011)

### 15.5 Survey data / recruit series

For the time being, pollack is caught in the IBTS survey only in small numbers; however, in the Skagerrak-Kattegat the cpue was much higher in the 1970s. They are distributed mainly over the northern North Sea (along the Norwegian Deep) and into the Skagerrak-Kattegat. Time series of abundance (average number per hour) in the IBTS are shown for Subarea 4 and Division 3.a separately, for quarter 1 (from 1983 onwards) and quarter 3 (from 1996 onwards) (Figure 15.2). The catches are small, and rather irregular, and no clear patterns emerge in 3 and 4.

### 15.5.1 Biological sampling

There has been some collection of length data in Subarea 4 and Division 3.a by Norway in the most recent years. Preliminary analysis of this data indicates that length ranges of pollack caught in gill net fisheries differ with meshsize and location. The majority of fish caught in western Norwegian fjords had a size range of $60-80 \mathrm{~cm}$ (Figure 15.3) compared to $50-70 \mathrm{~cm}$ in the Skagerrak (Figure 15.4).

### 15.5.2 Analysis of stock trends

In previous years the study by Cardinale et al. (2012), which analysed the spatial distribution and stock trends for the period 1906-2007, based on IBTS Q1 and commercial catches, was used to assess the stock for Division 3.a (Skagerrak and Kattegat) and it was found that there had a been large decline in stock size from approximately 1960 to 2000. However, during routine IBTS surveys in Subarea 4 and Subarea 3, pollack catches seem rather irregular and with no clear pattern. A spatial analysis of Norwegian fisheries data from 2013, showing total Pollack catches by ICES rectangle, indicates that the surveys do not cover the geographic distribution of the species adequately in both Subarea 4 and subdivision 3.a (Figures 15.5 and 15.6). The surveys may therefore not be very well suited for monitoring this species as trends in standardised CPUE likely are not a reliable indicator for the status of the stock. However, if the stock increases, it is arguably expected that present trawl survey (e.g. IBTS) would be able to detect such a stock trend in a consistent manner (Cardinale et al., 2012).

### 15.5.3 Data requirements

In order to get a better understanding of growth and maturity WGNEW recommended that the collection of otoliths and maturity should be continued during these surveys
for a few years. WGNSSK recommends also that the Norwegian biological data from commercial catches should be processed.

### 15.6 References

Cardinale, M., H. Svedäng, V. Bartolino, L. Maiorano, M. Casini and H. Linderholm, 2012. Spatial and temporal depletion of haddock and pollack during the last century in the KattegatSkagerrak. J. Appl. Ichthyol. 28(2): 200-208

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Heino, M., Svåsand, T., Nordeide, J. T., Otterå, H., 2012. Seasonal dynamics of growth and mortality suggest contrasting population structure and ecology for cod, pollack, and saithe in a Norwegian fjord. - ICES Journal of Marine Science 69: 537-546

Vølstad, J. H., Korsbrekke, K., Nedreaas, K. H., Nilsen, M., Nilsson, G. N., Pennington, M., Subbey, S.,Wienerroither, R., 2011. Probability-based surveying using self-sampling to estimate catch and effort in Norway's coastal tourist fishery. ICES Journal of Marine Science. 68: 1785-1791

Table 15.1. Pollack in Subarea 4 and Division 3.a. Landings (tonnes) by country as officially reported to ICES 1977-2015.

|  | ICES DIVISION 3.A |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Belgium | Denmark | Germany | Netherl. | Norway | Sweden | UK | Official Total |
| 1977 | 10 | 1764 | 4 | 3 | 449 | 706 |  | 2936 |
| 1978 | 1 | 2077 | 4 |  | 556 | 794 |  | 3432 |
| 1979 | 13 | 1898 | <0.5 |  | 824 | 1066 |  | 3801 |
| 1980 | 13 | 1860 |  |  | 987 | 1584 | <0.5 | 4444 |
| 1981 | 5 | 1661 |  |  | 839 | 1187 | 1 | 3693 |
| 1982 | 1 | 1272 |  |  | 575 | 417 | $<0.5$ | 2265 |
| 1983 | 2 | 972 |  |  | 438 | 288 |  | 1700 |
| 1984 | 2 | 930 | <0.5 |  | 371 | 276 |  | 1579 |
| 1985 | - | 824 | <0.5 |  | 350 | 356 |  | 1530 |
| 1986 | 4 | 759 | $<0.5$ |  | 374 | 271 |  | 1408 |
| 1987 | 6 | 665 |  |  | 342 | 246 |  | 1259 |
| 1988 | 4 | 494 |  |  | 350 | 136 |  | 984 |
| 1989 | 3 | 554 |  |  | 313 | 152 |  | 1022 |
| 1990 | 8 | 1842 | <0.5 |  | 246 | 253 |  | 2349 |
| 1991 | 2 | 1824 |  |  | 324 | 281 |  | 2431 |
| 1992 | 8 | 1228 |  |  | 391 | 320 |  | 1947 |
| 1993 | 6 | 1130 | 1 |  | 364 | 442 |  | 1943 |
| 1994 | 5 | 645 | <0.5 |  | 276 | 238 |  | 1164 |
| 1995 | 10 | 497 |  |  | 322 | 271 |  | 1100 |
| 1996 |  | 680 |  |  | 309 | 273 |  | 1262 |
| 1997 |  | 364 | <0.5 |  | 302 | 178 |  | 844 |
| 1998 |  | 299 |  |  | 330 | 105 |  | 734 |
| 1999 |  | 192 |  |  | 342 | 88 |  | 622 |
| 2000 |  | 199 |  |  | 268 | 33 |  | 500 |
| 2001 |  | 201 | 1 |  | 253 | 46 |  | 501 |
| 2002 |  | 228 | 3 |  | 202 | 44 |  | 477 |
| 2003 |  | 168 | 3 | 1 | 236 | 17 |  | 425 |
| 2004 |  | 140 | 2 | 4 | 179 | 34 |  | 359 |
| 2005 |  | 160 | 5 | 7 | 173 | 153 |  | 498 |
| 2006 |  | 103 | 10 | 3 | 178 | 36 |  | 330 |
| 2007 |  | 172 | 9 |  | 245 | 38 |  | 464 |
| 2008 |  | 166 | 5 |  | 247 | 33 |  | 451 |
| 2009 |  | 208 | 7 |  | 220 | 38 |  | 473 |
| 2010 |  | 313 | 8 | 1 | 195 | 35 |  | 552 |
| 2011 |  | 193 | 7 |  | 168 | 28 |  | 395 |
| 2012 |  | 200 | 7 |  | 171 | 37 |  | 414 |
| 2013 |  | 210 | 3 |  | 172 | 35 |  | 420 |
| 2014 |  | 191 | 5 | 1 | 156 | 30 |  | 383 |
| 2015 |  | 190 | $14$ | 1 | 138 | 48 |  | 389* |

[^6]|  | ICES Subarea 4 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Belgium | Denmark | Faeroes | France | Germany | Netherl. | Norway | Poland | Sweden | UK | Total |
| 1977 | 121 | 275 |  | 75 | 142 | 38 | 419 | 9 | 0 | 442 | 1521 |
| 1978 | 102 | 249 |  | 98 | 154 | 21 | 492 | 2 | 0 | 471 | 1589 |
| 1979 | 62 | 333 |  | 72 | 64 | 8 | 563 | 11 | 31 | 429 | 1573 |
| 1980 | 82 | 407 |  | 66 | 58 | 2 | 1095 |  | 38 | 355 | 2103 |
| 1981 | 59 | 500 |  | 173 | 21 | 2 | 1261 |  | 12 | 362 | 2390 |
| 1982 | 46 | 431 |  | 59 | 40 | 1 | 1169 | 33 | 23 | 270 | 2072 |
| 1983 | 58 | 481 |  | 79 | 44 | 1 | 1081 |  | 57 | 300 | 2101 |
| 1984 | 52 | 402 |  | 108 | 37 | 0 | 880 | 2 | 106 | 315 | 1902 |
| 1985 | 14 | 308 |  | 69 | 23 | 0 | 686 |  | 51 | 363 | 1514 |
| 1986 | 44 | 550 |  | 45 | 21 | 0 | 602 |  | 67 | 362 | 1691 |
| 1987 | 21 | 427 |  | 988 | 21 | 0 | 471 |  | 40 | 290 | 2258 |
| 1988 | 32 | 432 |  | 367 | 30 | 10 | 560 |  | 20 | 296 | 1747 |
| 1989 | 31 | 273 |  | 0 | 21 | 4 | 568 |  | 37 | 269 | 1203 |
| 1990 | 44 | 924 |  | 0 | 34 | 3 | 651 |  | 126 | 366 | 2148 |
| 1991 | 31 | 1464 |  | 0 | 48 | 4 | 887 |  | 153 | 684 | 3271 |
| 1992 | 49 | 794 |  | 18 | 59 | 7 | 1051 |  | 141 | 1310 | 3429 |
| 1993 | 46 | 1161 |  | 8 | 161 | 19 | 1429 |  | 217 | 1561 | 4602 |
| 1994 | 42 | 635 |  | 12 | 55 | 14 | 845 |  | 113 | 872 | 2588 |
| 1995 | 56 | 532 | 1 | 7 | 84 | 18 | 1203 |  | 175 | 1525 | 3601 |
| 1996 | 13 | 366 |  | 4 | 99 | 13 | 909 |  | 82 | 945 | 2431 |
| 1997 | 20 | 272 | 1 | 1 | 115 | 11 | 733 |  | 82 | 1185 | 2420 |
| 1998 | 21 | 265 |  | 7 | 44 | 5 | 567 |  | 75 | 780 | 1764 |
| 1999 | 21 | 288 |  | 0 | 62 | 5 | 768 |  | 72 | 636 | 1852 |
| 2000 | 45 | 291 |  | 24 | 38 | 5 | 880 |  | 91 | 877 | 2251 |
| 2001 | 36 | 156 |  | 6 | 40 | 1 | 860 |  | 63 | 809 | 1971 |
| 2002 | 27 | 234 |  | 6 | 112 | 0 | 879 |  | 68 | 711 | 2037 |
| 2003 | 13 | 191 |  | 9 | 82 | 1 | 971 |  | 36 | 837 | 2140 |
| 2004 | 28 | 162 |  | 5 | 57 | 0 | 517 |  | 16 | 612 | 1397 |
| 2005 | 26 | 173 |  | 3 | 128 | 3 | 511 |  | 46 | 477 | 1367 |
| 2006 | 18 | 152 |  | 4 | 80 | 1 | 545 |  | 12 | 587 | 1399 |
| 2007 | 18 | 192 |  | 130 | 137 | 2 | 754 |  | 43 | 905 | 2181 |
| 2008 | 15 | 150 |  | 129 | 114 | 1 | 840 |  | 46 | 999 | 2294 |
| 2009 | 13 | 121 | 2 | 6 | 50 | 1 | 668 |  | 32 | 658 | 1551 |
| 2010 | 12 | 163 |  | 10 | 129 | 0 | 599 |  | 32 | 540 | 1485 |
| 2011 | 12 | 106 | 0 | 10 | 67 | 0 | 580 | 0 | 35 | 489 | 1299 |
| 2012 | 17 | 123 | 0 | 3 | 102 | 1 | 433 |  | 42 | 443 | 1164 |
| 2013 | 17 | 128 | 0 | 2 | 66 | 4 | 371 | 0 | 29 | 463 | 1080 |
| 2014 | 24 | 121 |  | 32 | 145 | 1 | 476 |  | 40 | 377 | 1215 |
| 2015 | 19 | 183 |  | 2 | 237 | 2 | 473 |  | 50 | 625 | 1591* |

* Preliminary

Table 15.2. Pollack in Subarea 4 and Division 3.a. Catches (tonnes) by country as estimated by the Working Group 2013-2015.

|  | ICES Division 3.A |  |  |
| :--- | :--- | :--- | :--- |
|  | 2013 | 2014 | 2015 |
| Denmark | 214 | 192 | 192 |
| Germany | 11 | 6 | 35 |
| Netherlands | $<0.5$ | 0 | 0 |
| Norway | 174 | 156 | 138 |
| Sweden | 36 | 30 | 46 |
| ICES Total | 435 | 384 | 413 |
| Official Total | 420 | 383 | $389^{*}$ |
| Diff Ices-Off | 15 | 1 | 24 |

* Preliminary

| ICES SUBAREA 4 |  |  |  |
| :--- | :--- | :--- | :--- |
|  | 2013 | 2014 | 2015 |
| Belgium | 17 | 24 | 20 |
| Denmark | 150 | 122 | 183 |
| France | 2 | 32 | 2 |
| Germany | 59 | 145 | 216 |
| Netherland. | 3 | 1 | 2 |
| Norway | 379 | 481 | 466 |
| Sweden | 29 | 41 | 50 |
| UK | 456 | 377 | 626 |
| Ices Total | 1103 | 1227 | 1567 |
| Official Total | 1080 | 1215 | $1591^{*}$ |
| Diff Ices-Off | 23 | 12 | -22 |

* Preliminary

Table 15.3. Pollack in Subarea 4 and Division 3.a. Discards (tonnes) by country estimated by the Working Group, 2015.

| ICES DIvision 3.A |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Belgium | Denmark | Germany | Netherl. | Norway | Sweden | UK | Total |
| 2013 |  | 1.949 | 0.139 | 1.795 | 1.528 | 5.41 |  |  |
| 2014 | 0.62 | 0.008 | 0.441 | 0.473 | 1.54 |  |  |  |
| 2015 | 2.026 | 0.385 | 0.667 | 0.094 | 3.17 |  |  |  |


|  |  |  |  |  | ICES | UBAREA 4 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Belgium | Denmark | Faeroes | France | Germany | Netherl. | Norway | Poland | Sweden | UK | Total |
| 2013 | 0.111 | 22.785 |  | 0.050 | 0.229 | 1.320 | 7.967 |  | 0.662 | 8.923 | 42.05 |
| 2014 | 0.181 | 0.973 |  | 0.241 | 0.154 | 0.009 | 5.200 |  | 0.309 | 4.461 | 12.16 |
| 2015 |  | 0.069 |  | 0.005 | 0.075 | 0.001 | 0.691 |  | 0.090 | 1.59 | 2.52 |

[^7]

Figure 15.1. Pollack. Total landings of pollack from 2007-2015 in Division 3.a and Subarea 4 as officially reported to ICES.


Figure 15.2. Time series of catches of pollack from 1983-2015 in ICES divison 3.a (top graph) and Subarea 4 in the IBTS Q1 (red) and Q3 (blue) surveys, shown as numbers caught per hour with the GOV-trawl. Data from Datras.


Figure 15.3 Length distributions of pollack sampled by the Norwegian reference fleet in the years 2010 (top left panel), 2011 (top right panel), 2012 (bottom left panel) and 2013 (bottom right panel), Area 3.a. The data is aggregated for gillnets with a 63 mm meshsize.


Figure 15.4 Length distributions of pollack sampled by the Norwegian reference fleet in the years 2010 (top left panel), 2011 (top right panel), 2012 (bottom left panel) and 2013 (bottom right panel), Area 4. The data is aggregated for gillnets with a 70 mm mesh size.


Figure 15.5 Distribution of total pollack catches (Norwegian landings) for 2013 aggregated by fishing gear (bottom trawls, set nets, shrimp trawls), and pollack catches from IBTS surveys in 2012 (grey) and 2013 (green).


Figure 15.6 Pollack catches from IBTS surveys in 2013 (green) and 2014.

## 16 Grey gurnard (Eutrigla gurnardus) in Subarea 4 and Divisions 7.d and 3.a (North Sea, Eastern English Channel, Skagerrak and Kattegat)

### 16.1 General

Grey gurnard (Eutrigla gurnardus) was assessed in the Working Group on the Assessment of New MoU Species (ICES, 2014) until 2014. Biennal advice was given for the years 2015 and 2016. In 2015 the stock was assessed by the WGNSSK. For this stock several survey data were available. Available official landings data are incomplete or were not reported specifically for grey gurnard in the past. Only survey trends were used as a stock indicator (mature biomass index IBTS Q1). Based on the updated assessment the advised total catch should not be more than 8813 t for 2017 and 2018. This corresponds to landings not higher than 1763 t if the average discard rate $(80 \%)$ does not change.

### 16.1.1 Biology and ecosystem aspects

Grey gurnard occurs in the Eastern Atlantic from Iceland, Norway, southern Baltic, and North Sea to southern Morocco and Madeira. It is also found in the Mediterranean and Black Seas. In the North Sea and in Skagerrak/Kattegat, grey gurnard is an abundant demersal species. In the North Sea, the species may form dense semi-pelagic aggregations in winter to the northwest of the Dogger Bank, whereas in summer it is more widely distributed. The species is less abundant in the Channel, the Celtic Sea and in the Bay of Biscay.

Spawning takes place in spring and summer. There do not seem to be clear nursery areas. Grey gurnard can reach a maximum length of approximately 50 cm .

Grey gurnard is considered a predator on a number of commercially important demersal stocks (cod, whiting, haddock, sandeel, and Norway pout) in the North Sea (de Gee \& Kikkert, 1993). The steep increase in abundance of grey gurnard has led to an increase in mortality especially of North Sea cod (age-0) and whiting (age-0 and age-1) in recent years (ICES, 2011). The multi species model SMS estimated that grey gurnard can cause up to $50 \%$ of the predation mortality on 0 -group cod and whiting. Therefore, the abundance and distribution pattern of grey gurnard and its prey size preferences are highly relevant from an ecological point of view (Floeter and Temming, 2005; Kempf et al. 2013).

### 16.1.2 Stock ID and possible assessment areas

No studies are known of the stock ID of grey gurnard. In a pragmatic approach for advisory purposes and in order to facilitate addressing ecosystem considerations, the population is currently split among 3 Ecoregions: North Sea including 7.d, Celtic Seas and South European Atlantic. This proposal should be discussed considering the low levels of catches reported in recent years in Celtic Seas and South European Atlantic (ICES, 2011; ICES, 2012).

### 16.1.3 Management regulations

There is no minimum landing size for this species and there is no TAC.

### 16.2 Fisheries data

### 16.2.1 Historical landings

Historically, grey gurnard is taken as a by-catch species in mixed demersal fisheries for flatfish and roundfish. Grey gurnard from the North Sea is mainly landed for human consumption purposes. A high amount of grey gurnard is landed as industrial bycatch in the Danish fishery for sandeel and sprat (MIS_MIS_0_0_0_IBC). However, the market is limited and the largest part of the catch is discarded (see also stock Annex). Owing to the low commercial value of this species, landings data do not reflect the actual catches.

In the past, gurnards were often not sorted by species when landed and were reported as one generic category of "gurnards". Further, catch statistics are incomplete for some years, e.g. the Netherlands did not report gurnards during the years 1984-1999. In recent years, the official statistics seem to improve gradually. However, some countries continue to report "gurnards" landings and do not provide information on grey gurnard separately (e.g. Germany) or the data imported into InterCatch are based on a gurnard mix (e.g. UK England).

Since the early 1980s specific landings data for grey gurnard are available from the official catch statistics. Before that, these data occurred only sporadically in the statistics. Most of gurnard catches are taken in area IV and to a much lesser extent in areas 7.d and 3.a (Fig. 16.2.1.1.-16.2.2.3.; Table 16.1.1.-16.1.3.). Exceptionally high annual landings were reported during the late 1980s to early 1990s with a maximum of 46598 t in 1987 (Fig. 16.2.1.2; Table 16.1.2.) because of Danish landings for reduction purposes. After this peak, the Danish landings dropped again to a low level. Recent international landings for the last 5 years have been low ranging between 388 to 558 t per year. The average official landings for the period 2000-2014 was at 449 t . Data from 1950 to 2005 were taken from the "ICES catch statistics 1950 to 2010". Data from 2006 to 2014 were taken from the "ICES catch statistics 2006 to 2014". Data for 2015 were taken from the preliminary catch statistics.

### 16.2.2 InterCatch data

InterCatch contains now data for the years 2012-2015. Similar as for 2014, the largest amount of landings in 2015 was reported by Denmark for the MIS_MIS_0_0_0_IBC metier (1188t) which is mainly industrial fishery for sand eel and sprat. These landings are not included in the official landings which is the main reason for the large discrepancy between the official landings and the InterCatch estimate. Considerable amounts of landings were also reported by Scotland (297t, OTB_DEF_>=120_0_0_all) and Norway (171t, OTB_DEF_>=120_0_0_all). For all other metiers the landings were below 100 $t$ (Fig. 16.2.2.1). For all countries the amount of discards exceeded by far the amount of landings, with the exception of Denmark (Fig. 16.2.2.3). The largest amounts of discards were reported for the Scottish OTB_DEF_>=120_0_0_all (926 t), Scottish OTB_CRU_70-99_0_0_all (743 t) and the Dutch TBB_DEF_70-99_0_0_all (803 t) metiers. Not all countries reported grey gurnard discards (Norway, Belgium, Germany, France). The largest amount of discards was estimated for the OTB_DEF_>120_0_0_all fleet ( $\sim 1638 \mathrm{t}$ reported plus raised discards). The total catch estimated with Inter Catch for the year 2015 was 7316 t from which 1999 t were landings $(27 \%)$ and 5290 t estimated discards ( $73 \%$ of total catch). In total The Netherlands take the largest proportion of the total catch with a high amount of discards, followed
by Scotland. In 2015 landings were $15 \%$ lower compared to 2014 and total catches were 9\% lower.

The estimate of the previous year InterCatch landings and discards were revised this year by including German data and splitting UK England data to get an estimate for grey gurnard only. Germany does not report officially grey gurnard data separately, but rather reports a combined group of gurnards. Thus, it was not possible to upload German data into InterCatch. The uploaded InterCatch data from UK England were also based on a gurnard mix for which a ratio obtained by survey data was applied. This latter approach will lead to a bias because gurnard landings are usually dominated by tub gurnards (Chelidonichthys lucerna) while the largest part of grey gurnard is discarded. In order to estimate the grey gurnard proportion of these data the grey gurnard proportion of all gurnards from Dutch and Belgian official landings was used. This resulted in an average of $20 \%$ grey gurnards in landings for the three recent years (2014-2012). This ratio was then applied to the German and UK England data. Table 16.2.2.1 displays the change in total catch due to this correction.

### 16.2.3 Other information on Discards

In Table 16.2.3.1. the numbers per hour of discarded grey gurnard in Dutch bottomtrawl fisheries in North Sea and Eastern Channel are shown for 2006-2012 (Uhlmann et al., 2013). The rates are highly variable depending on the specific métiers, with highest values observed for the SSC_DEF métiers. German discard data from an observer programme indicate that the proportion of discarded gurnard in German demersal trawl fisheries ranges between $76.6 \%$ and $93.0 \%$ (Ulleweit et al., 2010).

### 16.3 Survey data / recruit series

For the North Sea and Skagerrak/Kattegat, data are available from the International Bottom Trawl survey. The IBTS-Q1 and IBTS-Q3 can provide information on distribution and the length composition of the catches. Grey gurnard occurs throughout the North Sea and Skagerrak/Kattegat. During winter, grey gurnards are concentrated to the northwest of the Dogger Bank at depths of 50-100 m, while densities are lower off the Danish coast, in the German Bight and eastern part of the Southern Bight (Figure 16.3.1. and 16.3.2.). The distribution pattern changes substantially in spring, when the whole area south of $56^{\circ} \mathrm{N}$ becomes densely populated and the high concentrations in the central North Sea disappear until the next winter (Daan et al. 1990).

The nearly absence of grey gurnard in the southern North Sea during winter and the marked shift in the centre of distribution between winter and summer suggests a preference for higher water temperatures (Hertling, 1924; Daan et al. 1990).
During winter, grey gurnard occasionally form dense aggregations just above the sea bed (or even in midwater, especially during night time) which may result in extremely large catches. Within one survey, these large hauls may account for $70 \%$ or more of the total catch of the species. Bottom temperatures in high-density areas usually range from 8 to $13^{\circ} \mathrm{C}$ (Sahrhage, 1964).

### 16.4 Biological sampling

Individual biological data for this species are still scarce (see also the stock annex). In the North Sea, individual data have been collected sporadically during some years of the IBTS-Q1 and IBTS-Q3 survey. An ALK from collected otoliths has shown that the age span of grey gurnard collected in Q1 is large (age 2 to age 14), but not many individuals were aged.

Available data on grey gurnard individual weights and maturity were analysed in order to estimate a mature biomass index (Figure 16.4.1). A maturity ogive based on all the available grey gurnard maturity data from IBTS Q1 was used to calculate the mature biomass index. The obtained maturity ogive shows that above 19.5 cm more than $90 \%$ of all the individuals can be considered mature (Figure 16.4.1.a). The corresponding Lmat $50 \%$ value was 15.6 cm . Proportion mature at length was calculated by the obtained model PropMat $=0.995 /\left(1+\exp \left(-1^{*}(\right.\right.$ LngtClass-15.611 $\left.\left.) / 2.073\right)\right)$. The obtained weight-length relation was Weight $=\left(0.007{ }^{*}\right.$ LngtClass $\left.{ }^{\wedge} 3.062\right)$.

The available age and maturity data suggest that grey gurnard is early maturing in North Sea and a certain proportion of fish at age 1 are mature.

### 16.5 Analysis of stock trends / assessment

Information from landings is very poor, due to poor reporting (gurnard species are not always identified in the data, and probably also misreporting has occurred) and also because the low value of the species leads to massive discarding.

The status of the populations in the Ecoregions which cover the Northern European Shelf is not known but some indications of trend are delivered by the survey series available.

To analyse stock trends a mature biomass index was calculated applying a length weight relationship and a maturity ogive which were obtained from all data available.

According to van Heesen and Daan (1996), outliers were excluded from the IBTS-Q1 time series since grey gurnards tend to form dense concentrations during winter. Outliers were defined as hauls which accounted for more than $90 \%$ of the total gurnard weight caught in the respective year. However, such extreme outliers were only identified in the time period before 1983 which is not displayed here. The time series of mature biomass index of grey gurnard of the IBTS-Q1 survey has shown a strong increase pattern from the beginning of 90's (Figure 16.5.1.; Table 16.5.1). Since then it was fluctuating on a high level. The mature biomass index for the IBTS-Q3 does not show this pronounced increasing trend but the 2014 value was the highest observed in the time series ever. In 2015 the IBTS-Q3 index dropped quite sharply again. In general lower biomass and abundance values were observed for the IBTS-Q3 survey time series. Compared to the North Sea/Skagerrak (area IV/3.a) the mature biomass values recorded by the Channel Ground Fish Survey (CGFS) in the Eastern Channel (area 7.d) were extremely low (not shown in this report). No trend could be detected in the CGFS index. Therefore, the advice for grey gurnard in area 4, 3.a and 7.d should be based on the IBTS survey, which covers by far the largest part of the stock.

### 16.5.1 DLS 3.2. approach

Grey gurnard was defined as a category 3 species following the ICES guidelines for data limited stocks (ICES, 2012). Consequently, the basis of the advice was a trend based assessment applying method 3.2 of the guidelines for data limited stocks:

$$
C_{y+1}=C_{y-1}\left(\frac{\sum_{i=y-x}^{y=y-1} I_{i} / x}{\sum_{i=y-z}^{y-x-1} I_{i} /(z-x)}\right)
$$

Where $C_{y+1}$ is the advised catch for the next year, $C_{y-1}$ should be the average catch of the last three years, and $I$ is the stock index. By default $x=2$ and $z=5$. A mature biomass index in kg per hour was estimated from the IBTS Q1 survey, because this survey covers most of the distribution area of grey gurnard. The stock size indicator (mature
biomass kg /hour) in the last two years (2015-2016) is $10 \%$ higher than the average of the three previous years (2012-2014). This results in an advised total catch of no more than 8813 t and no more than 1763 t landings given that the average discard rate of the previous three years does not change.

### 16.6 Data requirements

For management purposes, information should be available on catches and landings. Traditionally the quality of landings data has been poor for this species because in the past often only landings of "gurnards" were reported which is still the case for some countries today (e.g. Germany, UK England). Further, this species is highly discarded and discard data are only available for the recent years (2012-2015).

Given the high level of discarding, observation at sea under DCF is the main source of information to better estimate the total catches.

For a better understanding of this species an increase in our knowledge of biological parameters is required. In the context of ecosystem considerations, it would be useful to obtain more information on age composition of the stock and its diet composition.

From the information presented here, it can be concluded that grey gurnard is currently of very limited commercial interest.

### 16.7 References

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Table 16.2.2.1 InterCatch revision of landings, discards and total catch.

| Year | old Catch (WGNSSK2015) | updated Catch (WGNSSK2016) | Change |
| :--- | :--- | :--- | :--- |
| 2012 | 8345 | 7262 | $-13 \%$ |
| 2013 | 10230 | 8710 | $-15 \%$ |
| 2014 | 8596 | 8009 | $-7 \%$ |
| 2015 | 8343 | 7316 | $-6 \%$ |

Table 16.2.3.1 Grey gurnard. Discards per hour of grey gurnard by different metiers in the Netherlands 2006-2012.

| Métier | TBB_DEF | TBB_DEF* | TBB_DEF | SSC_DEF | SSC_DEF | OTB_MCD | OTB_DEF | OTB_DEF |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mesh | 70-99 | 70-99 | 100-119 | 100-119 | >120 | 70-99 | 70-99 | 100-119 |
| 2006 | 68.3 |  |  |  |  |  |  |  |
| 2007 | 60.2 |  |  |  |  |  |  |  |
| 2008 | 34.3 |  |  |  |  |  |  |  |
| 2009 | 55 | 17 | 37 |  |  | 111 | 77 | 15 |
| 2010 | 81 | 10 | 109 |  |  | 47 | 52 | 110 |
| 2011 | 61 | 27 | 10 | NA | 119 | 27 | 55 | 70 |
| 2012 | 41 | 24 | 30 | 317 | 307 | 110 | 75 | 12 |

Table 16.5.1.1. Summary of the assessment

| Year | Official landings | ICES LANdings | ICES <br> CATCHES | ICES DISCARDS | DISCARD RATE | Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | 589 |  |  |  |  | 5.00 |
| 1984 | 265 |  |  |  |  | 14.31 |
| 1985 | 301 |  |  |  |  | 3.75 |
| 1986 | 326 |  |  |  |  | 9.42 |
| 1987 | 44422 |  |  |  |  | 4.61 |
| 1988 | 37445 |  |  |  |  | 2.61 |
| 1989 | 26470 |  |  |  |  | 6.85 |
| 1990 | 22303 |  |  |  |  | 9.04 |
| 1991 | 14741 |  |  |  |  | 8.74 |
| 1992 | 8365 |  |  |  |  | 9.76 |
| 1993 | 1060 | - |  |  |  | 11.19 |
| 1994 | 254 |  |  |  |  | 10.61 |
| 1995 | 211 |  |  |  |  | 11.73 |
| 1996 | 301 |  |  |  |  | 18.71 |
| 1997 | 253 |  |  |  |  | 25.75 |
| 1998 | 145 |  |  |  |  | 21.45 |
| 1999 | 254 |  |  |  |  | 45.24 |
| 2000 | 661 |  |  |  |  | 25.83 |
| 2001 | 690 |  |  |  |  | 20.34 |
| 2002 | 499 |  |  |  |  | 24.85 |
| 2003 | 525 |  |  |  |  | 20.35 |
| 2004 | 452 |  |  |  |  | 21.29 |
| 2005 | 378 |  |  |  |  | 23.97 |
| 2006 | 267 |  |  |  |  | 22.24 |
| 2007 | 279 |  |  |  |  | 25.42 |
| 2008 | 273 |  |  |  |  | 24.68 |
| 2009 | 285 |  |  |  |  | 20.18 |
| 2010 | 388 |  |  |  |  | 30.87 |
| 2011 | 440 |  |  |  |  | 29.86 |
| 2012 | 632 | 904 | 7262 | 6358 | 0.88 | 32.47 |
| 2013 | 526 | 975 | 8710 | 7735 | 0.89 | 25.47 |
| 2014 | 499 | 1761 | 8009 | 6248 | 0.78 | 25.63 |
| 2015 | 777 | 2026 | 7316 | 5290 | 0.72 | 28.96 |
| 2016 |  |  |  |  |  | 32.29 |

Table 16.1.1. Official grey gurnard landings in area 3.a.

| Year | BE | DK | NL | NO | SE | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0 | 0 | 0 | 0 | 36 | 36 |
| 1981 | 0 | 0 | 0 | 0 | 46 | 46 |
| 1982 | 0 | 86 | 0 | 0 | 43 | 129 |
| 1983 | 0 | 29 | 0 | 0 | 7 | 36 |
| 1984 | 0 | 62 | 0 | 0 | 6 | 68 |
| 1985 | 0 | 3 | 0 | 0 | 9 | 12 |
| 1986 | 0 | 6 | 0 | 0 | 10 | 16 |
| 1987 | 1 | 13 | 0 | 0 | 6 | 20 |
| 1988 | 0 | 59 | 0 | 0 | 2 | 61 |
| 1989 | 0 | 19 | 0 | 0 | 4 | 23 |
| 1990 | 0 | 34 | 0 | 0 | 3 | 37 |
| 1991 | 0 | 25 | 0 | 0 | 5 | 30 |
| 1992 | 0 | 22 | 0 | 0 | 10 | 32 |
| 1993 | 0 | 18 | 0 | 0 | 9 | 27 |
| 1994 | 0 | 12 | 0 | 0 | 12 | 24 |
| 1995 | 0 | 10 | 0 | 0 | 5 | 15 |
| 1996 | 0 | 18 | 0 | 0 | 3 | 21 |
| 1997 | 0 | 13 | 0 | 0 | 5 | 18 |
| 1998 | 0 | 27 | 0 | 0 | 8 | 35 |
| 1999 | 0 | 23 | 0 | 0 | 5 | 28 |
| 2000 | 0 | 32 | 0 | 0 | 5 | 37 |
| 2001 | 0 | 30 | 0 | 0 | 3 | 33 |
| 2002 | 0 | 18 | 0 | 0 | 1 | 19 |
| 2003 | 0 | 32 | 0 | 0 | 1 | 33 |
| 2004 | 0 | 24 | 2 | 0 | 2 | 28 |
| 2005 | 0 | 21 | 4 | 0 | 1 | 26 |
| 2006 | 0 | 19 | 0 | 0 | 2 | 21 |
| 2007 | 0 | 21 | 1 | 0 | 3 | 25 |
| 2008 | 0 | 24 | 0 | 0 | 5 | 29 |
| 2009 | 0 | 15 | 0 | 0 | 3 | 18 |
| 2010 | 0 | 10 | 1 | 0 | 2 | 13 |
| 2011 | 0 | 5 | 0 | 0 | 1 | 6 |
| 2012 | 0 | 5 | 0 | 0 | 1 | 6 |
| 2013 | 0 | 5 | 0 | 0 | 1 | 6 |
| 2014 | 0 | 3 | 0 | 0 | 1 | 4 |
| 2015 | 0 | 4 | 0 | 1 | 2 | 7 |

Table 16.1.2. Official grey gurnard landings in area IV.

| Year | BE | DK | FR | NL | NO | SE | UK | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0 | 0 | 43 | 0 | 0 | 0 | 0 | 43 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 100 | 0 | 0 | 0 | 0 | 100 |
| 1983 | 0 | 0 | 64 | 0 | 0 | 0 | 0 | 64 |
| 1984 | 0 | 0 | 71 | 0 | 0 | 0 | 0 | 71 |
| 1985 | 88 | 0 | 85 | 0 | 0 | 0 | 0 | 173 |
| 1986 | 0 | 27 | 66 | 0 | 0 | 0 | 0 | 93 |
| 1987 | 63 | 44205 | 56 | 0 | 0 | 0 | 0 | 44324 |
| 1988 | 72 | 36887 | 43 | 0 | 0 | 0 | 22 | 37024 |
| 1989 | 73 | 26230 | 45 | 0 | 0 | 0 | 0 | 26348 |
| 1990 | 85 | 22041 | 42 | 0 | 0 | 0 | 0 | 22168 |
| 1991 | 70 | 14514 | 28 | 0 | 0 | 0 | 0 | 14612 |
| 1992 | 98 | 8113 | 21 | 0 | 0 | 0 | 10 | 8242 |
| 1993 | 106 | 822 | 27 | 0 | 0 | 0 | 24 | 979 |
| 1994 | 63 | 87 | 21 | 0 | 0 | 0 | 22 | 193 |
| 1995 | 43 | 63 | 26 | 0 | 0 | 0 | 21 | 153 |
| 1996 | 108 | 52 | 18 | 0 | 0 | 0 | 54 | 232 |
| 1997 | 49 | 23 | 22 | 0 | 0 | 0 | 57 | 151 |
| 1998 | 33 | 29 | 13 | 0 | 0 | 0 | 0 | 75 |
| 1999 | 35 | 63 | 0 | 0 | 0 | 127 | 0 | 225 |
| 2000 | 28 | 63 | 5 | 452 | 0 | 0 | 0 | 548 |
| 2001 | 22 | 258 | 20 | 277 | 0 | 1 | 33 | 611 |
| 2002 | 23 | 45 | 10 | 285 | 0 | 1 | 29 | 393 |
| 2003 | 16 | 60 | 5 | 307 | 0 | 6 | 26 | 420 |
| 2004 | 21 | 59 | 6 | 264 | 0 | 3 | 23 | 376 |
| 2005 | 16 | 52 | 5 | 213 | 0 | 8 | 22 | 316 |
| 2006 | 10 | 46 | 2 | 133 | 2 | 0 | 7 | 200 |
| 2007 | 11 | 16 | 4 | 155 | 5 | 0 | 14 | 205 |
| 2008 | 8 | 24 | 2 | 104 | 5 | 3 | 12 | 158 |
| 2009 | 15 | 6 | 2 | 154 | 1 | 1 | 22 | 201 |
| 2010 | 14 | 8 | 10 | 218 | 1 | 0 | 13 | 264 |
| 2011 | 26 | 6 | 7 | 263 | 1 | 0 | 31 | 334 |
| 2012 | 49 | 3 | 4 | 467 | 2 | 0 | 77 | 602 |
| 2013 | 30 | 4 | 2 | 268 | 34 | 0 | 131 | 469 |
| 2014 | 35 | 4 | 3 | 252 | 56 | 0 | 128 | 478 |
| 2015 | 20 | 7 | 2 | 209 | 172 | 4 | 345 | 760 |

Table 16.1.3. Official grey gurnard landings in area 7.d.


Table 16.5.1. Grey gurnard mature biomass indices (kg/hour) from IBTS Q1 and IBTS Q3.

| Year | IBTS Q1 | IBTS Q3 |
| :---: | :---: | :---: |
| 1983 | 5.00 |  |
| 1984 | 14.31 |  |
| 1985 | 3.75 |  |
| 1986 | 9.42 |  |
| 1987 | 4.61 |  |
| 1988 | 2.61 |  |
| 1989 | 6.85 |  |
| 1990 | 9.04 |  |
| 1991 | 8.74 | 2.62 |
| 1992 | 9.76 | 5.98 |
| 1993 | 11.19 | 3.79 |
| 1994 | 10.61 | 4.92 |
| 1995 | 11.73 | 5.22 |
| 1996 | 18.71 | 9.26 |
| 1997 | 25.75 | 7.45 |
| 1998 | 21.45 | 12.41 |
| 1999 | 45.24 | 13.44 |
| 2000 | 25.83 | 9.40 |
| 2001 | 20.34 | 11.15 |
| 2002 | 24.85 | 7.81 |
| 2003 | 20.35 | 8.36 |
| 2004 | 21.29 | 3.98 |
| 2005 | 23.97 | 3.28 |
| 2006 | 22.24 | 3.52 |
| 2007 | 25.42 | 4.10 |
| 2008 | 24.68 | 7.00 |
| 2009 | 20.18 | 6.15 |
| 2010 | 30.87 | 5.16 |
| 2011 | 29.86 | 10.41 |
| 2012 | 32.47 | 5.78 |
| 2013 | 25.47 | 7.08 |
| 2014 | 25.63 | 12.14 |
| 2015 | 28.96 |  |
| 2016 | 32.29 |  |



Figure 16.2.1.1: Gurnards. Official landings of grey gurnard in area 3.a, 3b and 3c 1985-2015 (a), official landings of grey gurnards by country only in area 3.a 2006-2015 (b).


Figure 16.2.1.2: Gurnards. Official landings of grey gurnard in area IV for Denmark only(a) , official landings of grey gurnards by country in area IV since 1993 (b).


Figure 16.2.1.3: Gurnards. Official landings of grey gurnard in area 7.d and 7e (a), official landings of grey gurnards by country only in area $7 . \mathrm{d}$ by country since 2006 (b).


Figure 16.2.2.1 Inter Catch. Grey gurnard landings in 2015 by metier and country as uploaded to InterCatch. Panel (a) and (b) showing the same data, but y-axis in panel (b) is downscaled to better visualize details.


Figure 16.2.2.2 Inter Catch. Grey gurnard discards in 2015 by metier and country. Reported discards panel (a), raised discards panel (b). Legend valid for both panels.


Figure 16.3.1: Grey gurnard. Spatial distribution of grey gurnard from IBTS-Q1 survey in area IV and 3.a.


Figure 16.3.2: Grey gurnard. Spatial distribution of grey gurnard from IBTS-Q3 survey in area IV and 3.a.

## Catches



Stock size indicator


Figure 16.5.1.1: Estimated total catches and landings (upper panel) and stock indicator (lower panel).


Figure 16.4.1. (a) Maturity ogive of Grey gurnard sampled during IBTS Q1 surveys ( $\mathrm{n}=1501$ ), (b) length weight relationship of Grey gurnard sampled during IBTS Q1 surveys.


Figure 16.5.1. IBTS Q1 and IBTS Q3 grey gurnard mature biomass index.

## 17 Striped red mullet (Mullus surmuletus) in Subarea 4 and Divisions 7.d and 3.a (North Sea, Eastern English Channel, Skagerrak and Kattegat)

### 17.1 General

Striped red mullet has been benchmarked in 2015 (ICES 2015).
The main issues addressed during the benchmark were the quantity and representativeness of the observational data. Analyses suggested the extrapolation of the assessment results from the eastern English Channel to the southern North Sea had merit. It was less clear whether the assessment was valid for the other areas within the stock region, because the fishery catches were small and data were sparse.
The conclusion of the benchmark were, that the agreed stock assessment seemed reasonable given the available information and that it could be used for providing fisheries advice under the ICES Stock Category 3 framework. Ecosystem aspects
Striped red mullet (Mullus surmuletus) is a benthic species. Young fish are distributed in coastal areas, while adults have a more offshore distribution. Benzinou et al, (2013) conducted stock identification studies based on otolith and fish shape in European waters and showed that striped red mullet can be geographically divided into two units: Western Unit (Subareas 6 and 8, and Divisions 7.a-c, 7.e-k, and 9.a) and Northern Unit (Subarea 4 (North Sea) and Divisions 7.d (Eastern English Channel) and 3.a (Skager-rak-Kattegat))..
In the English Channel, this species matures at approximately 16 cm .
Juveniles are found in waters of low salinity, while adults are found at high salinity. Striped red mullet prefers sandy sediments.
Adult red mullet feed on small crustaceans, annelid worms and molluscs, using their chin barbels to detect prey and search the mud.

### 17.2 Fisheries

Historically, France has taken most of the landings with a targeted fishery for striped red mullet ( $>90 \%$ of landings). This French fishery targeting striped red mullet is conducted by bottom trawlers using a mesh size of 70-99 mm in the eastern English Channel and in the southern North Sea.

The eastern English Channel and southern North Sea areas are also fished by trawlers of various types targeting a variety of species. Striped red mullet might be a bycatch in these fisheries.

From 2000 a Dutch targeted fishery, using flyshooters, and a UK fisheries have also developed. Landings are shared by these three fleets in the latter years.

### 17.3 ICES advice

Advice for 2016 and 2017
ICES advises that when precautionary approach is applied, catches should be no more than 552 tonnes in each of the years 2016 and 2017.

All catch are assumed to be landed. Selectivity in the fishery should be improved to avoid fishing on juvenile recruits and to protect the strong 2014 year class.

### 17.4 Management

No specific management objectives are known to ICES. There is no TAC for this species.

There is no minimum landing size for this species.
Demersal fisheries in the area are mixed fisheries, with many stocks exploited together in various combinations in the various fisheries. In these cases, management advice must consider both the state of individual stocks and their simultaneous exploitation in demersal fisheries. Stocks in the poorest condition, particularly those which suffer from reduced reproductive capacity, become the overriding concern for the management of mixed fisheries, where these stocks are exploited either as a targeted species or as a bycatch.

### 17.5 Data available

### 17.5.1 Catch

Official landings data are shown by country in Table 17.5.1.1 and by area in Table 17.5.1.2. There is no indication of discard of striped red mullet. All catches are assumed to be landed. Table 17.5.1.3 presents total official landings and ICES estimates over the period 2006-2015 as well as the predicted catch corresponding to advice. In 2015 53\% of the catches were made using demersal seines and $30 \%$ using demersal trawls.

Total landings were provided under the ICES InterCatch format for the period 20032013 during the benchmark. However only France provided age composition for the period 2006-2013. 2014 and 2015 landings were provided under the ICES InterCatch format. Figure 17.5.1.1 shows that only landings from France in the Eastern Channel (representing around $30 \%$ of the total landings) were provided in 2014 and 2015 with an age structure. Figure 17.5.1.2 shows that IC data and official landings are consistent over years and countries.

Age composition of the landing were provided under the ICES InterCatch format for the period 2006-2013 for the benchmark in 2015 but only for the Fench Otter Trawlers in the Eastern Channel. All other fleets were raised using this stratum. Figure 17.5.1.3 and table 17.5.1.1 show the age structure of the landings provided in 2015.
Prior to 2009, no landings of age 0 were observed. Most of the landings are made on age 1.

### 17.5.2 Weight at age

Mean weight at age were computed as described in the stock annex and are presented in figures 17.5.2.1 and 17.5.2.2 and table 17.5.2.1.

Weights at age in the landings show a slight decrease for the oldest ages. However sampling intensity for these ages is very low due to the low number of fishes in the catches. Stock weight do not show this slight decrease of age 3 and $4+$ but as for landings weight, the sampling is very low due to the low number of fishes in the landings.

### 17.5.3 Maturity and natural mortality

Information about maturity per age class is given with the table included in this section. At an age of one year more than 50 percent of the striped red mullet are mature.

| AGE | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ |  | $\mathbf{5}$ | $\mathbf{6}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Maturity | 0 | 0.54 | 0.65 | 1 | 1 | 1 | 1 |  |

As defined during WKNSEA (ICES 2015), natural mortality was derived from Gislason first estimator (Gislason et al ., 2010) leading, as expected for this species, to high natural mortality for the youngest ages (see table included below).

|  | AGE | M_GISLASON |
| :--- | :--- | :--- |
| 0 | 1.426 |  |
| 1 | 0.6641 |  |
| 2 | 0.4888 |  |
| 3 | 0.4164 |  |
| 4 | 0.3616 |  |
| 5 | 0.3275 |  |
| 6 | 0.3421 |  |

### 17.5.4 Survey data

The Channel Ground Fish Survey (CGFS) and the IBTS-Q3 surveys were estimated to be good indicators of the population trends as they cover the spatial distribution of this stock. However none of them have an exhaustive coverage of the spatial distribution.
In 2015, a change in the research vessel used for the CGFS was realised. The consequences of these changes were assessed via an intercalibraton in 2014 and some analysis of the catch data (see WGIBTS report). It appeared that for red mullet indices seem to be used without correcting factor.
Only CGFS survey allowed deriving age structured indices. Internal consistencies of the survey (Fig. 17.5.4.1) show reasonable consistencies between age 1 and 4 .

The age composition of the catches made during CGFS is presented in Figure 17.5.4.2.

### 17.6 Trend based assessment

As agreed during WKNSEA (ICES 2015), the assessment model was used for trend as the SSB estimated by the model was considered to be a more reliable indicator of stock status than the direct use of survey indices.

The settings used are described on the following table.

| Setting/Data | VALUES/SOURCE |
| :--- | :--- |
| Catch at age | Landings (since 2004, ages 0-4+) InterCatch |
|  | Discards are assumed neglictible |
| Tuning indices | FR CGFS (since 2004 ages 0-4+) |
| Plus group | 4 |
| First tuning year | 2004 |
| Fishing mortality | $\sim$ s(year, k=5) + factor(age) |
| Survey catchability | $\sim$ factor(age) |
| Recruitment | $\sim$ factor(year) |

Results from the assessment are presented in figures 17.6.1. Log residuals of the model are presented in Figure 17.3.6.2 and observed and predicted catches in figures 17.6.3 and 17.6.4.

As observed during WKNSEA, there is still a relatively high uncertainty in this assessment but the SSB is still at a very low level and the recruitment seems to be the highest observed during this limited time series. The slight increase in SSB in mostly due to the few age 1 fishes left in the population. Trends show a very low level of biomass and a very high fishing mortality. Most of the catches rely only on the recruitment (age 0) and age 1 fishes.

### 17.7 Conclusions drawn from analyses

The very good recruitment observed in 2014 was confirmed by the catches in 2015 and the remaining age 1 seen in 2015 during CGFS. There is no TAC on that species so the advice was not followed and the catches overshot the advice for 2015 (4487 Tonnes against 460 Tonnes in the advice).
Basis for the advice:
For the previous reason and the poor recruitment observed in 2015, the advice was not reopened in 2016:
Striped red mullet in Subarea 4 and Divisions 7.d and 3.a. For stocks in ICES data categories 3-6, one catch option is possible. This is highlighted in bold.

| Indicator (2013-4) : SSB | 914 |  |
| :--- | :--- | :--- |
| Indicator (2010-2012) : SSB | 621 |  |
| Indicator ratio | 1.47 |  |
| Recent advice catch | 460 |  |
| Recent advice catch * indicator ratio | 676 |  |
| Uncertainty cap applied | Yes | 552 |
| Precautionary buffer applied | Yes | 2013 |

A good recruitment has been observed in the different surveys and landings have increased in 2014. However, the increase in landings and in indices only rely on the recruitment (age 0) and age 1.

Instead of using the average catches of the last three years, recent advice catch was taken as a basis for the calculation because ICES advice has never been implemented despite a substantial over exploitation of the stock.
The uncertainty cap has to be applied leading to an increase in TAC by 20\%

### 17.8 Sources:

Benzinoua, A., Carbinia, S., Nasreddinea, K., Elleboode, R., Mahé K., Discriminating stocks of striped red mullet (Mullus surmuletus) in the Northwest European seas using three automatic shape classification methods

ICES 2015. Report of the Benchmark Workshop on North Sea Stocks (WKNSEA), 2-6 February 2015. ICES CM 2015/ACOM:32

Table 17.5.1.1 Striped red mullet in Subarea 4 and Divisions 7.d and 3.a. Official and ICES landings by country (tonnes).

| .Year | Belgium | Denmark | France | Netherlands | UK | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 0 | 0 | 140 | 0 | 0 | 140 |
| 1976 | 0 | 0 | 156 | 3 | 1 | 160 |
| 1977 | 0 | 0 | 279 | 12 | 1 | 292 |
| 1978 | 0 | 0 | 207 | 25 | 3 | 235 |
| 1979 | 0 | 0 | 212 | 32 | 11 | 255 |
| 1980 | 0 | 0 | 86 | 25 | 4 | 115 |
| 1981 | 0 | 0 | 44 | 19 | 1 | 64 |
| 1982 | 0 | 0 | 32 | 18 | 2 | 54 |
| 1983 | 0 | 0 | 232 | 15 | 1 | 248 |
| 1984 | 0 | 0 | 204 | 0 | 3 | 207 |
| 1985 | 0 | 0 | 135 | 0 | 4 | 140 |
| 1986 | 0 | 0 | 84 | 0 | 3 | 88 |
| 1987 | 0 | 1 | 40 | 0 | 3 | 46 |
| 1988 | 0 | 1 | 35 | 0 | 4 | 41 |
| 1989 | 0 | 0 | 37 | 0 | 5 | 42 |
| 1990 | 0 | 0 | 524 | 0 | 13 | 537 |
| 1991 | 0 | 0 | 208 | 0 | 11 | 219 |
| 1992 | 0 | 0 | 458 | 0 | 17 | 475 |
| 1993 | 0 | 0 | 576 | 0 | 21 | 597 |
| 1994 | 0 | 0 | 362 | 0 | 18 | 380 |
| 1995 | 0 | 0 | 2537 | 0 | 69 | 2606 |
| 1996 | 0 | 2 | 2039 | 2 | 44 | 2087 |
| 1997 | 0 | 2 | 856 | 0 | 61 | 919 |
| 1998 | 0 | 2 | 2966 | 0 | 117 | 3085 |
| 19991) | 0 | 4 | NA | 0 | 103 | 107 |
| 2000 | 0 | 4 | 3201 | 464 | 133 | 3802 |
| 2001 | 0 | 10 | 1789 | 915 | 183 | 2897 |
| 2002 | 0 | 24 | 1658 | 560 | 141 | 2383 |
| 2003 | 28 | 0 | 3256 | 626 | 177 | 4087 |
| 2004 | 31 | 0 | 4137 | 1148 | 129 | 5445 |
| 2005 | 29 | 0 | 1918 | 914 | 136 | 2997 |
| 2006 | 126 | 0 | 1030 | 293 | 116 | 1451 |
| 2007 | 13 | 0 | 3475 | 906 | 292 | 4686 |
| 2008 | 15 | 0 | 3250 | 873 | 606 | 4744 |
| 2009 | 14 | 0 | 736 | 562 | 428 | 1740 |
| 2010 | 62 | 0 | 879 | 567 | 466 | 1974 |
| 2011 | 83 | 0 | 650 | 540 | 338 | 1611 |
| 2012 | 39 | 0 | 155 | 367 | 187 | 748 |
| 2013 | 33 | 0 | 112 | 180 | 42 | 367 |
| 2014 | 71 |  | 720 | 700 | 242 | 1732 |
| 2015 | 211 |  | 1598 | 1997 | 356 | 4162 |

${ }^{1)}$ No data reported by France in 1999.
${ }^{2)}$ ICES estimates.

Table 17.2.1.2 Striped red mullet in Subarea 4 and Divisions 7.d and 3.a. Official landings by area (tonnes). Note: Most of the Subarea 4 catches are made in Division 4.c.

| Year |  | 4 |  | 3.A |  | 7.D | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 0 |  | 0 |  | 140 |  | 140 |
| 1976 | 4 |  | 0 |  | 156 |  | 160 |
| 1977 | 19 |  | 0 |  | 273 |  | 292 |
| 1978 | 30 |  | 0 |  | 205 |  | 235 |
| 1979 | 49 |  | 0 |  | 206 |  | 255 |
| 1980 | 29 |  | 0 |  | 86 |  | 115 |
| 1981 | 20 |  | 0 |  | 44 |  | 64 |
| 1982 | 21 |  | 0 |  | 33 |  | 54 |
| 1983 | 41 |  | 0 |  | 207 |  | 248 |
| 1984 | 22 |  | 0 |  | 185 |  | 207 |
| 1985 | 10 |  | 0 |  | 130 |  | 140 |
| 1986 | 6 |  | 0 |  | 82 |  | 88 |
| 1987 | 7 |  | 0 |  | 38 |  | 46 |
| 1988 | 7 |  | 0 |  | 33 |  | 41 |
| 1989 | 5 |  | 0 |  | 37 |  | 42 |
| 1990 | 33 |  | 0 |  | 504 |  | 537 |
| 1991 | 26 |  | 0 |  | 193 |  | 219 |
| 1992 | 60 |  | 0 |  | 415 |  | 475 |
| 1993 | 126 |  | 0 |  | 471 |  | 597 |
| 1994 | 116 |  | 0 |  | 264 |  | 380 |
| 1995 | 1054 |  | 0 |  | 1552 |  | 2606 |
| 1996 | 528 |  | 0 |  | 1559 |  | 2087 |
| 1997 | 278 |  | 0 |  | 641 |  | 919 |
| 1998 | 778 |  | 0 |  | 2307 |  | 3085 |
| 19991) | 70 |  | 0 |  | 37 |  | 107 |
| 2000 | 1764 |  | 0 |  | 2038 |  | 3802 |
| 2001 | 1600 |  | 0 |  | 1297 |  | 2897 |
| 2002 | 1234 |  | 0 |  | 1149 |  | 2383 |
| 2003 | 1618 |  | 0 |  | 2469 |  | 4087 |
| 2004 | 1820 |  | 0 |  | 3625 |  | 5445 |
| 2005 | 1404 |  | 0 |  | 1593 |  | 2997 |
| 2006 | 338 |  | 0 |  | 1113 |  | 1451 |
| 2007 | 787 |  | 0 |  | 3899 |  | 4686 |
| 2008 | 946 |  | 0 |  | 3798 |  | 4744 |
| 2009 | 471 |  | 0 |  | 1269 |  | 1740 |
| 2010 | 359 |  | 0 |  | 1615 |  | 1974 |
| 2011 | 307 |  | 0 |  | 1304 |  | 1611 |
| 2012 | 196 |  | 0 |  | 552 |  | 748 |
| 2013 | 99 |  | 0 |  | 268 |  | 367 |
| 2014 | 263 |  | 0 |  | 1469 |  | 1732 |
| 2015 | 770 |  |  |  | 3392 |  | 4162 |

[^8]Table 17.3. 1.3 Striped red mullet in Subarea 4 and Divisions 7.d and 3.a. History of ICES advice, the agreed TAC, and ICES estimates of landings.

| Year | ICES Advice | Predicted catch CORRESP. TO ADVICE | Official landings | ICES Estimates |
| :---: | :---: | :---: | :---: | :---: |
| 2006 |  | - | 1451 | 1483 |
| 2007 |  | - | 4686 | 4610 |
| 2008 |  | - | 4744 | 2066 |
| 2009 |  | - | 1740 | 1518 |
| 2010 |  | - | 1974 | 1920 |
| 2011 |  | - | 1611 | 1512 |
| 2012 | No increase in catch | - | 748 | 725 |
| 2013 | No increase in catches (average 2009-2010) | $<1700$ | 367 | 409 |
| 2014 | Reduce catches by $36 \%$ compared to 2012 | < 460 | 1732 | 1717 |
| 2015 | No new advice, same as for 2014 | < 460 | 4162 | 4487 |
| 2016 | Precautionary approach | <552 |  |  |
| 2017 | Precautionary approach | <552 |  |  |

Weights in tonnes.

Table 17.5. 1.1 Striped red mullet landing numbers at age (thousands).

| AGE | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0 | 0 | 0 | 0 | 0 | 55 | 14734 | 0 | 6 | 1384 | 10124 | 1832 |
| 1 | 43375 | 16606 | 3912 | 37013 | 1323 | 16259 | 15203 | 9317 | 1335 | 2771 | 10790 | 37485 |
| 2 | 1839 | 2455 | 2332 | 1124 | 10518 | 1319 | 674 | 1454 | 1244 | 467 | 1329 | 6310 |
| 3 | 947 | 263 | 1679 | 553 | 1255 | 662 | 142 | 639 | 1477 | 289 | 14 | 19 |
| 4 | 187 | 256 | 188 | 127 | 537 | 102 | 102 | 80 | 183 | 0 | 29 | 36 |

Table 17.5. 2.1 Striped red mullet stock weights (kg).

| AGE | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0 | 0 | 0 | 0 | 0 | 0.046 | 0.042 | 0 | 0.02 | 0.02 | 0.029 | 0.038 |
| 1 | 0.09 | 0.105 | 0.15 | 0.107 | 0.096 | 0.07 | 0.077 | 0.05 | 0.09 | 0.06 | 0.093 | 0.1 |
| 2 | 0.222 | 0.172 | 0.19 | 0.313 | 0.139 | 0.16 | 0.112 | 0.15 | 0.17 | 0.12 | 0.144 | 0.114 |
| 3 | 0.27 | 0.3 | 0.24 | 0.422 | 0.226 | 0.177 | 0.24 | 0 | 0.25 | 0.12 | 0.259 | 0.37 |
| 4 | 0.569 | 0.411 | 0.37 | 0.506 | 0.361 | 0.423 | 0.209 | 0.02 | 0.23 | 0 | 0.309 | 0.2 |



Figure 17.5. 1 Striped red mullet in Subarea 4 and Divisions 7.d ICES landings by country (percentage over the total area).


Figure 17.5. 2 Striped red mullet in Subarea 4 landings (comparison between IC data, red line) and official catch statistics (black and blue for provisional)


Figure 17.5.1.3 Striped red mullet age structure as provided in 2014 for the Feench


Figure 17.5.1.3 Striped red mullet age structure (in numbers) as provided in the landings


Figure 17.5.2.1 Weight at age in the stock


Figure 17.5.2.2 Weight at age in the landings


Lower right panels show the Coefficient of Determination $\left(r^{2}\right)$

Figure 17.5.4.1 CGFS internal consistencies

CGFS, index 2016


Figure 17.5.4.2 CGFS catch age composition
red mullet VIId IV IIIa


Figure 17.6.1 CGFS internal consistencies
log residuals of catch and abundance indices


Figure 17.6.2
Log residuals of the assessment


Figure 17.6.3 observed (pink) and estimated (blue) catch number-at-age


Figure 17.6.4 observed (pink) and estimated (blue) indices at age

## 18 Turbot (Scophthalmus maximus) in Subarea 4 (North Sea)

This report presents the stock assessment carried out for turbot (Scophthalmus maxima) in Subarea 4 in 2015. Following an inter-benchmark procedure for this stock, a new assessment model (SAM) was used since 2015. More details on the data used, assumptions made and the assessment model settings can be found in the stock annex.

Turbot was assessed at WGNSSK for the first time in 2013. At IBPNEW 2012 (ICES, 2012) an assessment model was developed for turbot in Subarea 4. While the assessment model developed for turbot represented a big step forward, there was still much work to be done to fine tune the model settings to improve the reliability and consistency of outputs. This lead to an Inter-benchmark procedure in 2014-15 (IBPturbot; ICES 2015). Following IBPturbot, a new technique for modelling weights at age has been developed, a plusgroup (10+) was added to the assessment and the assessment methodology was changed.

At WGNSSK 2013 (ICES, 2013a) it was decided to categorise turbot as a 'Category 2' stock under the ICES data-limited stocks (DLS) framework (ICES, 2013b). Category 2 stocks have quantitative assessments that are treated as indicative of trends rather than absolute values. The assessment model adapted during IBPturbot is proposed as a Category 1 assessment, but numerous data quality issues and poor diagnostics mean that the assessment of this stock is still highly uncertain. After an external review highlighting the many issues with the assessment, the advice drafting group North Sea finally decided to treat the assessment output (SSB) as indicative of trends and to use this as the basis for deriving advice under category 3 of the ICES DLS approach.

### 18.1 General

### 18.1.1 Biology and ecosystem aspects

Turbot is broadly distributed from Iceland in the North, along the European coastline, to the Mediterranean and Adriatic Sea in the south. In general, turbot is a rather sedentary species, but there are some indications of migratory patterns. For example in the North Sea, migrations from the nursery grounds in the south-eastern part to more northerly areas have been recorded. IBPNEW (ICES, 2012) concluded that Turbot in the North Sea (Subarea 4) can be considered as a distinct stock for management purposes.

Turbot is typically found at a depth range of 10 to 70 m , on sandy, rocky or mixed bottoms and is one of the few marine fish species that inhabits brackish waters. It is a typical visual feeder and could be regarded as a top predator. Turbot feeds mainly on bottom living fishes (e.g. common gadoids, sandeels, gobies, sole, dab, dragonets, sea breams etc.) and small pelagic fish (e.g. herring, sprat, boarfish, sardine) but also, to a lesser extent, on larger crustaceans and bivalves. Despite its role as a top predator in the North Sea ecosystem, at present turbot is not included as a species in the WGSAM multispecies assessment (ICES, 2014a).

### 18.1.2 Fisheries

In the 1950s the UK was the biggest contributor to the landings ( $\sim 50 \%$ of the landings). In recent years most of the landings stem from the Netherlands ( $\sim 50-60 \%$ ). In most countries turbot is caught in mixed fisheries trawls, with most of the landings in the Netherlands coming from the 80 mm beam trawl fleet (BT2) fishing for sole and plaice.

In Denmark, the second largest contributor to the landings in recent times, there is a directed fishery for turbot using gillnets ( $\sim 10 \%$ of the total landings).
See the stock annex (section A.2) for more details.

### 18.1.3 ICES advice for 2016

The information in this section is taken from the ICES advice sheet 2015, section 6.3.54. This stock is managed under a biennial TAC (together with brill), but ICES has provided advice individually for each of these stocks.

## Advice for 2016:

ICES advises that when the precautionary approach is applied, catches should be no more than 1995 tonnes in each of the years 2016 and 2017. If discard rates do not change from 2014, this implies landings of no more than 1925 tonnes.

Management of turbot and brill under a combined species TAC prevents effective control of the single species exploitation rates and could lead to the overexploitation of either species.

### 18.1.4 Management

A combined EU TAC for turbot and brill is set for EU waters in areas IIa and 4. This TAC only applies to the EU fisheries. This management area (particularly the inclusion of area IIa) does not correspond to either of the stock areas defined by ICES for turbot and brill.

No specific management objectives or plans are known to ICES. Following IBPturbot precautionary reference points ( $\mathrm{B}_{\mathrm{lim}}$ and $\mathrm{B}_{\mathrm{pa}}$ ) have been proposed for the stock.

As a primarily bycatch species, regulations relating to effort restrictions for the primary metiers catching turbot (e.g. beam trawlers) are likely to impact on the stock. Fishing effort has been restricted for demersal fleets in a number of EC regulations (e.g. EC Council Regulation Nos. 2056/2001, 51/2006, 41/2007, and 40/2008).

The Dutch Producer Organisations have introduced a minimum landings size of 27 cm in 2015, and 30 cm in 2016 in order to maintain the landings within the national quota.
See the stock annex (section A.2) for more details.

### 18.2 Data used

To estimate the trends in abundance and exploitation over time, the assessment of the turbot stock requires three main types of data:
Catch data: estimates of removals of turbot by the fishery.
Survey data and commercial LPUE (landings per unit effort): indices of trends in population abundance over time from fisheries independent and fisheries dependent sources, respectively.
Biological data: estimates and/or assumptions on growth, maturation and natural mortality.
Since the assessment is age-based, data for the above is required for each age. See the stock annex (section B) for more details on the data used in the assessment, sources and historical values.

### 18.2.1 Catch data

The assessment model only uses catch numbers at age and does not utilise total landings (tonnage per year). InterCatch was used for the first time for the North Sea turbot stock at WGNSSK 2014, and has been used since.

Age structure of the landings for 2004-2015 has been estimated from Dutch samples accounting for auctions, quarters and market categories. Prior to 2014, all samples were taken from the 80 mm beam trawl fishery (TBB_DEF_70-99). In 2012, 11 samples were taken and a total of 596 fish were aged. In 2013, 8 samples were taken and a total of 426 fish were aged. In 2014, 9 samples ( 503 aged fish) were taken from the TBB_DEF_7099 metier and 2 samples ( 120 aged fish) were taken from otter trawl metier (OTB_DEF_70-99). In 2015, 14 samples (781 aged fish) were taken from the TBB_DEF_70-99 metier.

Figure 18.2.1 shows the metiers with numbers at age samples for the landings and the age distributions by season observed in these metiers. The only usable samples were those from the Dutch TBB_DEF_70-99 (beamtrawl) metier, and these samples were used to raise all the unsampled métiers in 2014.

Raising was done by quarter. The TBB_DEF_70-99 samples were evenly spread over the seasons,. All beam trawl fleets were raised using the age distributions calculated form the TBB samples. There are a wide variety of métiers that land turbot, including a significant amount from the Danish $>220 \mathrm{~mm}$ gillnet fleet. More samples of landings at age are required from other métiers to more accurately raise the total landings at age.

Figure 18.2.2 shows the trend in total landings over time and Figure 18.2.3 shows the breakdown of landings by country for 2014 (from the EuroStat database). Landing of turbot decreased during the 1990s and for the last ten years have been stable in the region of 3000 t . Over this time effort by the Dutch beam trawl fleet, which contributes the most of the landings, has decreased notably. Since turbot is primarily a bycatch species, this indicates that abundance of turbot has likely increased over this period.

Landings at age are presented in Table 18.2.1 and Figure 18.2.4. The 2005 yearclass shows up clearly in the landings data, but since then there have been no notably large year classes observed. Following a decrease in minimum market size for turbot in the Netherlands in 2002, there has been a notable increase in the amount of age 1 and 2 turbot landed, accounting for half of the catch in some years but this proportion has been decreasing in recent years due to some poor year classes in 2012 and 2013. Since turbot are only fully mature at age 4 , indicates a high proportion of immature fish in the landings. However, the last 5 years have also seen an increase in the proportion of age $5+$ fish in the landings compared to the five years prior to that, though still lower than observed in the 1970s and 1980s. This could reflect a reduction in F recently leading to an increasing proportion of older fish in the landings. However, since the catch data is raised using only the Dutch 80 mm TBB fleet, signals in catch at age data may not be accurate reflections of true removals from the population over time.

### 18.2.1.1 Discard data

The assessment of this stock assumes that discarding of catches for this stock is negligible. However, there was a sudden increase in the landing of age two turbot following the decrease in minimum market size in the Netherlands in 2002. Given that there was no known change in the fishing behaviour of the main fleets at this time, this could indicates that previously more age 2 fish must have been caught than were actually
landed. These were either discarded or, as a much sought after fish, kept by the fishermen for personal use. This would mean that the discards could be underestimated in the period up to 2002 relative to the period following this, potentially causing a bias in the assessment outputs. Alternatively, subsequent to the change in MLS, more targeting of small turbot may have occurred. Without a useable time series of discards before and after this change it is difficult to determine which of these explanations holds.

However, the impact on the final year estimates is likely to be small because with the reduction in minimum market size in 2002, the assumption of negligible discards probably holds for the last 10 years. Discard data were submitted to Intercatch by various nations. However, there is very limited age sampling of the discards. Very few fish were sampled in the discards of some of the Danish métiers ( $<10$ per métier, fewer than the number of ages in the assessment model), not enough to be used in the raising of international landings.

In 2016, most countries provided estimates of discards in 2015 to Intercatch. Out of the 2925 t that were landed, $2012 \mathrm{t}(86 \%)$ had associated reported discards (totalling 92 t ). When the rest of the discards were raised for the unsampled landings an extra 27 t of discards is estimated. In total this gives an estimate of $92+20=112 \mathrm{t}$ discarded, implying a discard rate of $3.7 \%$ in 2014. No useable age structure information was submitted for the discard estimates.

Overall, discard rates estimated from the available data suggest less than $5 \%$ of the catch is discarded and are therefore discarding is considered negligible.

### 18.2.2 Survey data and commercial LPUE

Two survey abundance indices, the Sole Net Survey (SNS) and the Beam Trawl Survey (BTS ISIS), and one commercial LPUE abundance index, the Dutch 80 mm beam trawl fleet (BT2), are used to tune the assessment (Table 18.2.2 and Figure 18.2.5). Prior to IBPturbot the Dutch BT2 LPUE index was used as an age-structured index of abundance. Following IBPturbot and WGNSSK 2015 it was decided to rather use this index and age-aggregated index of exploitable biomass. This was decided upon since the same catch at age data was used to raise the catch at age matrix and the Dutch BT2 index. It was felt that feeding the same data into the model for both the catch and the index would inherently bias the assessment in favour of this index since it follows very close the catch information used in the model. The new age-aggregated exploitable biomass per unit effort index is shown in Figure 18.2.5b.

All abundance indices indicate an increase in the number of fish aged 4 and older in late 2000 s compared to the past. An increase in the amount of older fish would indicate either strong recruitment or a decrease in mortality (e.g. fishing pressure) exerted on the stock. However, following 2010 the indices indicate a decrease in some of the older ages and there are no clear indications of strong year classes in the most recent years. In particular, estimates of numbers at age 2 are low in 2014. Both fisheries independent surveys however, show large numbers of age 1 and age 2 fish in 2015.

There is fairly close agreement between the three indices on the general trends in abundance at age, but the data are noisy from year to year. This can be seen in the low $\mathrm{R}^{2}$ values in the internal consistency correlations in the BTS_ISIS and SNS surveys (Figure 18.2.9). The SNS survey is particularly poor at picking up cohort signals, with low $\mathrm{R}^{2}$ values on the correlations between numbers at consecutive ages. Though all correlations between successive ages are positive, estimated numbers at age, particularly for the younger ages, fluctuate a lot from year to year. The BTS-ISIS is more internally
consistent for ages 3 and up. The almost non-existent relationship between the numbers estimated at age 1 and the numbers estimated at age 2 in the following year suggest that in future removing age 1 , and potentially age 2 , from this index may be appropriate. The internal consistency of the NL_BT2 LPUE index is significantly better, though the removal of age 1 from this index could also be considered. However, this index is no longer used as an age-structured index.
Noisy indices that are more indicative of general trends are best used in an assessment model that is able to smooth over the noise in the data. The SAM model used for this stock is able do this, but nevertheless inputting noisy data into the assessment will increase uncertainty in the outputs. By removing the age-structure from the NL BT2 LPUE index, the clearest cohort signals in the assessment of this stock are coming from the catch at age matrix.

### 18.2.3 Biological data

All biological data used in the assessment are presented in Table 18.2.3.

## Weight at age

Constant annual catch and stock weights at age (long term means of all available data) were previously used in the assessment because of large gaps in the time series of weight at age data for turbot in the North Sea (Figure 18.2.7). What data is available is also very noisy, due to low sample sizes for most ages. The data that are available, and trends in other flatfish species in the same areas suggest that there have been potentially significant changes in weight at age over time. At IBPturbot a method was developed to model the growth parameters over time, allowing smooth changes over the time series (see stock annex for full details). The results indicate an increase in weight at age from the start of the time series, peaking in the early 1990s. Since then weights at age have decreased again to slightly lower than the 1970s, and have been fairly stable in recent years.

## Maturity

At IBPNEW (ICES, 2012) turbot maturity data from the Netherlands was used to study some reproductive characteristics of turbot from the North Sea. A female maturity ogive constructed from derived from a General Linear Model fit using the maturity data from the recent time period was chosen for the stock.

## Natural mortality

There are currently no accepted estimates of turbot natural mortality over time. A number of alternative methods, using different estimates of growth parameters, were used to estimate the level of natural mortality by age for turbot in the North Sea at IBPNEW (ICES, 2012). Since turbot grows relatively fast compared to other flatfish species in the same areas, results indicate that natural mortality is higher. However, due to high variability for recorded values of $K$ (an estimated growth parameter) for turbot, it proved difficult to find agreement on natural mortality values. Hence, after performing assessment test runs, a constant value of $\mathrm{M}=0.2$ for all ages and years was chosen for this stock. This is twice the level used in the sole and plaice assessments in the North Sea.

### 18.3 Stock assessment model

Turbot in Subarea 4 was previously assessed using a custom designed age-structured assessment model. Following IBPturbot, a SAM model is now used for this stock. The
basic SAM set up was modified to allow the inclusions of an exploitable biomass index (see stock annex).

### 18.3.1 Model settings

The assessment model was conducted using the settings and configuration given below. Details of the assessment model can be found in the stock annex.

Assessment settings used in the final assessment

| Year | 2015 (IBPTURBOT PROPOSAL) |
| :---: | :---: |
| Model | SAM |
| First tuning year | 1975 |
| Last data year | 2014 |
| Ages | 1-10+ |
| Plus group | Yes |
| Stock weights at age | Von Bertalanffy growth curve with time varying Linf |
| Catch weights at age | Von Bertalanffy growth curve with time varying Linf |
| Total Landings | Not used |
| Landings at age | 1975-1978, 1981-1990, 1998, 2000-present |
| Discards | Not used (assumed 0) |
| Abundance indices | BTS-Isis 1985-2013 |
|  | SNS 1975-2002, 2004-2013 |
|  | NL-BT2 LPUE age-aggregated catchable biomass 2002-2014 |
| Catchability independent of age for ages $>=$ | 7 |

SAM configuration file (see stock annex for details)

```
# Min Age
1
# Max Age
10
# Max Age considered a plus group (0=No, 1=Yes)
1
# The following matrix describes the coupling of fishing mortality STATES
# Row represent Catch, Columns represent ages.
1 [lllllllllll
# Use correlated random walks for the fishing mortalities
# ( 0 = independent, 1 = correlation estimated, 2=AR1)
2
# Coupling of catchability PARAMETERS (Surveys)
# Row represent fleets ( SNS and BTS only; LPUE age-aggregated), Columns represent
ages.
\begin{tabular}{llllllllll}
1 & 2 & 3 & 4 & 5 & 6 & 6 & 0 & 0 & 0 \\
7 & 8 & 9 & 10 & 11 & 12 & 12 & 0 & 0 & 0
\end{tabular}
# Coupling of power law model EXPONENTS
(not used)
# Coupling of fishing mortality RW VARIANCES
1 1 2 
# Coupling of }\operatorname{log}\mathrm{ N RW VARIANCES
1 [lllllllll
# Coupling of OBSERVATION VARIANCES
# Row represent fleets (Catch, SNS, BTS), Columns represent ages.
\begin{tabular}{llllllllll}
1 & 2 & 3 & 3 & 3 & 3 & 3 & 3 & 4 & 4 \\
5 & 6 & 7 & 7 & 7 & 7 & 7 & 0 & 0 & 0 \\
8 & 9 & 10 & 10 & 10 & 10 & 10 & 0 & 0 & 0
\end{tabular}
# Stock recruitment model code ( 0=RW, 1=Ricker, 2=BH, ... more in time)
0
# Years in which catch data are to be scaled by an estimated parameter
(Catch not scaled)
# Define Fbar range
2 6
```


### 18.4 Assessment model results

Abundance at age and fishing mortality at age estimated by the assessment model are presented in Tables 18.4.1 and 18.4.2, respectively. Key stock and fishery metrics are given in Table 18.4.3 and plotted in Figure 18.4.1.

### 18.4.1 Status of the stock

Fishing mortality was estimated at 0.99 in 2015, a sharp increase from 2013 (0.74). This is well above the long term geometric mean (0.49). The SSB in 2014 was estimated to be 3469 t , decreasing sharply to 2569 t in 2015. Both years are lower than the long term geometric mean ( 7177 t ). The estimated recruitment (age 1) for 2015 is higher than the geometric mean of the time series. However, this estimate is based on very little data and is unlikely to be a reliable estimate.

### 18.4.2 Historic stock trends

Spawning stock biomass since 2000 has been at a low level compared to the period before this.

SSB peaked in the early 1980s, at a time when F was estimated to be at the lowest level of the time series. From the mid-1980s up until the early 2000s SSB declined gradually and F increased gradually. The lowest observed SSB was in 2005, SSB subsequently increased until 2010 and has remained between 3500 t and 400t since then. However, SSB at the start of 2015 is estimated to have declined sharply since 2014 to 2610 t , below $\mathrm{B}_{\mathrm{pa}}$. This is because recruitment in 2013 is estimated to be the lowest in the time series. Likewise, the 2014 and 2012 year classes are estimated to be poor.

Mean F peaked in 2003 at 0.93 , but then declined sharply to 0.46 in 2006 before gradually increasing again. This corresponds to a period of reduction in effort by the NL BT2 fleet and allowed SSB to gradually increase from its lowest observed level. The perceived increase in mean F (ages 2-6) in the early 2000s is the result of an increase in the amount of age 2 turbot landed following the change in MLS in the Netherlands. Since the mean F range covers ages $2-6$, this increase in F on age 2 fish lead to a sharp increase in mean F over this period.

There are no clear patterns in recruitment, though values are estimated at a slightly higher level, but with more uncertainty, during the years of missing landings at age data (1990s). Recent recruitment has been poor, with the exception of the most recent year class.

### 18.4.3 Retrospective assessments

The results of five retrospective assessments, run using the same model settings but removing one year of data from the end of the time series, are plotted in Figure 18.4.2. In most years F has been severely overestimated and SSB has been underestimated.. For SSB only one of the retrospective peels falls within the confidence bounds of the most recent assessment. For F, three retrospective peels are well above the confidence bounds of the latest assessment. Recruitment shows no clear retrospective pattern and all peels fall within the confidence bounds of the latest assessment.

The disagreement between retrospective assessments is large, and though not clearly biased, are still problematic. This is due to the quality of the data used by the assessment rather than the model itself.

### 18.5 Model diagnostics

Diagnostic tests are carried out on the assessment model fit and outputs to identify any irregularities or potential biases that should be taken into account when interpreting the assessment results. The diagnostics for the turbot SAM model are poor.
Table 18.5.1 shows the observation and process error estimates from the SAM model. The process error is not particularly large. The process error on numbers at age for age 1 is estimated to be three times as high as that for the other ages. This means that the dynamics between age 1 and age 2 are uncertain, potentially diminishing the meaning of the recruitment estimates (since year class strength as estimated at age 1 may not correlate very strongly with year class strength as estimated at age 2). The observation variance for almost all ages of the SNS and BTS-ISIS indices are very large. This is unsurprising given the poor internal consistency found in these indices. Observation variances on the catch are particularly high for age 1 (2.19), indicating that estimates for catches at age one are significantly down weighted by the model. The estimates for age 9 and the plusgroup are also high (0.61), while the observations variance for all the other ages are lower than those for the survey indices (with the exception of age 2 for the BTS-ISIS).

For the fishery there is a clear periods of negative or positive residuals for age 1 and the model has consistently estimated more age 2 catches in recent years than were raised. Ages $2-4$, which contribute to the vast majority of the landings, are estimated by the model to be most similar to the observed values. The increase in age 3+ fish observed in the 2014 catch at age results in negative residuals for most of these ages in 2014 (model predicts less than reported).

The residuals for the three tuning indices are presented Figure 18.5.3. There are some year effects present in the SNS index. In 2014 all residuals are negative, while in 2012 and 2013 most of the residuals were positive. The BTS-ISIS index has clear groupings of negative or positive residuals over certain ages and years. The NL BT2 exploitable biomass index has positive residuals until 2005 and negative residuals since then indicating that the index suggests a more rapid increase in biomass over this time period than the model. The fact that all indices (with the exception of age 2 in the BTS-ISIS) have negative residuals in the final year suggests that the indices are have a limited impact on the

To evaluate the impact of each of the tuning indices on the model fit, leave-one-out (LOO) runs were conducted (Figure 18.5.4). Excluding the SNS index leads to higher SSB and lower F in the last 7 years. SSB in particular is estimated to be well above the confidence bounds of the full assessment. During the period when no catch at age data are available (the 1990s), leaving out the SNS leads to lower SSB and higher F, while leaving out the BTS-ISIS leads to higher SSB and lower F. Leaving out the NL BT2 exploitable biomass index does a very negligible effect on SSB, F and recruitment estimation. This suggests that this index is down weighted strongly in the model, hence the very poor residual pattern for this index. Recruitment values do not change significantly is any of the indices are removed. The indicates that the catch at age matrix is the major source of data contributing to the models estimates of year class strength.

### 18.5.1 Data Limited Stocks (DLS) approach

Given the poor diagnostics of the assessment, an alternative approach to deriving advice for this stock is to use the ICES framework for category 3 stocks (ICES 2012a). An SSB index from the age based assessment treated as indicative of trends can be applied as the indicator of stock development. The perception of the stock has not changed with the inclusion of the new data.

### 18.6 Management considerations

There are a number of EC regulations that affect the flatfish fisheries in the North Sea, e.g. as a basis for setting the TAC, limiting effort, and minimum mesh size.

### 18.6.1 Effort regulations

Regulated effort restrictions in the EU were introduced in 2003 (annexes to the annual TAC regulations) for the protection of the North Sea cod stock. In addition, a long-term plan for the recovery of cod stocks was adopted in 2008 (EC regulation 1342/2008). In 2009, the effort management programme switched from a days-at-sea to a kW -day system (EC regulation 43/2009), in which different amounts of kW -days are allocated within each area by member state to different groups of vessels depending on gear and mesh size. Effort ceilings are updated annually. The most important fleet catching turbot in the North Sea, the Dutch 80mm beam trawl fleet (BT2), was affected by this regulation only once in 2009 but not since.

The overall fleet capacity and deployed effort of the North Sea beam trawl fleet has been substantially reduced since 1995, due to a number of reasons, including the above mentioned effort limitations for the recovery of the cod stock. In 2008, 25 vessels were decommissioned.

### 18.6.2 Technical measures

Turbot is mainly taken by beam trawlers in a mixed fishery directed at sole and plaice in the southern and central part of the North Sea. Technical measures (EC Council Regulation $1543 / 2000$ ) applicable to the mixed flatfish fishery affect the catching of turbot. The minimum mesh size of 80 mm in the beam trawl fishery selects sole at the minimum landing size ( 24 cm ). However, this mesh size is likely to catch immature turbot (age 1 and 2 fish). Mesh enlargement would reduce the catch of smaller turbot at the same time potential increasing the yield per recruit, but would also result in loss of marketable sole catches.

A closed area has been in operation since 1989 (the plaice box) and since 1995 this area has been closed in all quarters. The closed area applies to vessels using towed gears, but vessels smaller than 300 HP are exempted from the regulation. An additional technical measure concerning the fishing gear is the restriction of the aggregated beam length of beam trawlers to 24 m . In the 12 nautical mile zone and in the plaice box the maximum aggregated beam-length is 9 m .

### 18.6.3 Combined TAC

At present the EU provides a combined TAC for turbot and brill in the North Sea. It is unclear how the quantitative single species advice for turbot and the qualitative single species advice for brill can/will be used to formulate a combined TAC for these two stocks. In this situation, improving the brill assessment may be necessary in order to ensure efficient management of both of these stocks. Ideally, a combined TAC would not be used.

### 18.7 References

ICES, 2012. Report of the Inter-Benchmark Protocol on New Species (Turbot and Sea bass; IBPNew 2012), 1-5 October 2012, Copenhagen, Denmark. ICES CM 2012/ACOM: 45. 239pp.

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ICES. 2015. Report of the Inter-Benchmark Protocol for Turbot in Subarea IV (IBP Turbot), May 2015, By correspondence. ICES CM 2015/ACOM:XXX. Xxx pp.

Table 18.2.1 Turbot in Subarea 4: Landings at age estimates (abundance, thousands) used in the assessment.

|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1975 | 1.15 | 427.35 | 1012.35 | 239.35 | 108.35 | 124.55 | 90.35 | 47.25 | 42.05 | 146.55 |
| 1976 | 0.35 | 350.35 | 1346.35 | 392.35 | 114.35 | 76.25 | 57.75 | 50.55 | 38.55 | 174.15 |
| 1977 | 18.55 | 895.35 | 644.35 | 531.35 | 166.35 | 44.15 | 30.85 | 42.35 | 36.95 | 142.35 |
| 1978 | 0.35 | 1324.35 | 1273.35 | 309.35 | 268.35 | 76.35 | 37.95 | 29.35 | 20.75 | 65.05 |
| $1979-1980$ | NO DATA |  |  |  |  |  |  |  |  |  |
| 1981 | 0.35 | 299.35 | 755.35 | 532.35 | 458.35 | 175.35 | 67.35 | 35.35 | 40.35 | 32.35 |
| 1982 | 0.35 | 169.35 | 1046.35 | 267.35 | 167.35 | 292.35 | 98.35 | 49.35 | 41.35 | 65.35 |
| 1983 | 0.35 | 402.35 | 673.35 | 479.35 | 110.35 | 113.35 | 180.35 | 91.35 | 31.35 | 81.35 |
| 1984 | 0.35 | 1296.35 | 1223.35 | 311.35 | 157.35 | 60.35 | 57.35 | 74.35 | 51.35 | 70.35 |
| 1985 | 0.35 | 795.35 | 2415.35 | 654.35 | 179.35 | 109.35 | 26.35 | 38.35 | 48.35 | 74.35 |
| 1986 | 0.35 | 371.35 | 1470.35 | 697.35 | 183.35 | 67.35 | 29.35 | 16.35 | 18.35 | 90.35 |
| 1987 | 13.35 | 648.35 | 546.35 | 676.35 | 158.35 | 52.35 | 19.35 | 5.35 | 5.35 | 60.35 |
| 1988 | 36.35 | 1084.35 | 897.35 | 178.35 | 176.35 | 90.35 | 28.35 | 42.35 | 10.35 | 25.35 |
| 1989 | 0.35 | 594.35 | 1037.35 | 315.35 | 139.35 | 73.35 | 28.35 | 22.35 | 10.35 | 29.35 |
| 1990 | 43.35 | 957.35 | 1032.35 | 305.35 | 160.35 | 73.35 | 98.35 | 58.35 | 13.35 | 39.35 |
| $1991-1997$ | NO DATA |  |  |  |  |  |  |  |  |  |
| 1998 | 0.35 | 540.35 | 1158.35 | 476.35 | 97.35 | 39.65 | 11.65 | 10.45 | 1.26 | 8.35 |
| 1999 | NO DATA |  |  |  |  |  |  |  |  |  |
| 2000 | 4.85 | 255.35 | 938.35 | 270.35 | 315.35 | 145.05 | 116.45 | 51.65 | 59.14 | 72.7 |
| 2001 | 0.35 | 478.35 | 1642.35 | 357.35 | 64.35 | 75.85 | 55.45 | 65.05 | 21.93 | 61.49 |
| 2002 | 468.35 | 1283.35 | 1237.35 | 265.35 | 123.35 | 32.45 | 16.75 | 17.65 | 3.62 | 9.87 |
| 2003 | 267.35 | 2429.35 | 586.35 | 378.35 | 90.35 | 42.25 | 26.65 | 9.35 | 8.02 | 9.43 |
| 2004 | 491.15 | 2234.35 | 894.35 | 156.35 | 93.35 | 11.25 | 8.85 | 4.45 | 1.37 | 1.73 |
| 2005 | 291.45 | 1678.35 | 611.35 | 195.35 | 21.35 | 18.85 | 2.55 | 12.15 | 1.38 | 3.42 |
| 2006 | 706.05 | 1312.35 | 644.35 | 95.35 | 28.35 | 6.65 | 13.25 | 3.45 | 1.05 | 10.66 |
| 2007 | 80.25 | 2829.35 | 627.35 | 290.35 | 41.35 | 29.95 | 8.75 | 9.85 | 0.35 | 6.34 |
| 2008 | 184.85 | 1404.35 | 854.35 | 229.35 | 203.35 | 49.35 | 13.75 | 1.55 | 7.16 | 2.62 |
| 2009 | 117.25 | 1076.35 | 1005.35 | 434.35 | 92.35 | 26.25 | 11.75 | 8.15 | 2 | 9.71 |
| 2010 | 237.15 | 1193.35 | 328.35 | 263.35 | 146.35 | 75.25 | 26.35 | 6.35 | 4.83 | 6.14 |
| 2011 | 219.45 | 2017.35 | 626.35 | 115.35 | 143.35 | 80.35 | 33.85 | 16.65 | 3.72 | 4.84 |
| 2012 | 0.35 | 1949.35 | 793.35 | 272.35 | 43.35 | 65.25 | 74.55 | 13.65 | 6.88 | 4.71 |
| 2013 | 161.68 | 1480.55 | 1013.13 | 304.82 | 85.22 | 24.34 | 39.35 | 16.95 | 2.76 | 4.54 |
| 2014 | 94.56 | 548.41 | 620.09 | 461.38 | 162.79 | 78.15 | 24.22 | 21.31 | 11.21 | 32.21 |
| 2015 | 57.39 | 1911.06 | 531.05 | 347.21 | 354.88 | 111.12 | 34.10 | 12.59 | 11.60 | 17.554 |

Table 18.2.2 Turbot in Subarea 4: Relative abundance indices used in the assessment. SNS.

| SNS |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| start | 0.66 | end | 0.75 |  |  |  |  |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1975 | 92.7765 | 80.0575 | 19.9635 | 6.8025 | 2.5955 | 2.9235 | 0.6775 |
| 1976 | 43.9395 | 53.5585 | 12.8505 | 4.5435 | 2.2625 | 1.5405 | 0.4575 |
| 1977 | 406.4435 | 212.4895 | 43.5875 | 12.0325 | 8.4625 | 3.3085 | 1.7745 |
| 1978 | 27.5655 | 121.5145 | 49.0915 | 17.3695 | 7.5775 | 4.5425 | 2.3125 |
| 1979 | 14.5295 | 107.7745 | 45.9225 | 13.9525 | 8.0695 | 4.8685 | 2.4875 |
| 1980 | 109.9805 | 66.3845 | 24.5475 | 8.4455 | 3.7515 | 2.5315 | 1.6905 |
| 1981 | 23.6485 | 65.2125 | 23.7685 | 7.4655 | 8.0285 | 5.1455 | 4.2085 |
| 1982 | 87.7025 | 40.0155 | 7.5965 | 1.7605 | 1.2525 | 0.3035 | 0.7955 |
| 1983 | 151.5805 | 146.6065 | 27.7185 | 8.7335 | 2.7935 | 2.9135 | 1.1675 |
| 1984 | 88.7745 | 76.0395 | 24.9595 | 9.5415 | 5.1865 | 3.1565 | 1.3265 |
| 1985 | 42.6505 | 93.7175 | 21.9185 | 6.8335 | 3.9385 | 3.0565 | 0.7115 |
| 1986 | 24.1895 | 15.9245 | 4.9955 | 1.8845 | 2.7715 | 3.3135 | 1.2595 |
| 1987 | 62.2215 | 16.8645 | 3.2955 | 1.0885 | 0.2885 | 0.6675 | 0.4785 |
| 1988 | 166.6585 | 101.5035 | 17.9345 | 4.2315 | 1.1895 | 0.6835 | 0.6684 |
| 1989 | 62.5885 | 44.8695 | 13.2645 | 5.7775 | 3.4475 | 1.1415 | 0.7695 |
| 1990 | 231.0075 | 102.8355 | 22.3815 | 4.7845 | 2.2265 | 0.2685 | 2.7135 |
| 1991 | 37.5065 | 75.4595 | 21.0285 | 6.2545 | 2.6065 | 2.5775 | 0.6465 |
| 1992 | 249.0995 | 106.4985 | 33.3155 | 14.6285 | 7.5245 | 4.7545 | 0.7845 |
| 1993 | 146.6255 | 154.4875 | 33.8125 | 9.1675 | 3.8695 | 2.8315 | 0.8465 |
| 1994 | 94.9145 | 47.8945 | 16.5355 | 9.2765 | 4.5725 | 0.9505 | 1.3515 |
| 1995 | 189.3095 | 58.2185 | 5.2265 | 2.5675 | 0.8135 | 0.8495 | 0.6245 |
| 1996 | 80.0705 | 79.0695 | 17.8295 | 4.7915 | 1.1525 | 2.1045 | 0.8065 |
| 1997 | 31.3715 | 27.4255 | 9.3865 | 5.3905 | 4.1575 | 0.5675 | 1.6635 |
| 1998 | 53.3635 | 41.9375 | 9.8765 | 2.5165 | 1.6535 | 0.9105 | 0.5725 |
| 1999 | 156.4185 | 97.0725 | 28.3925 | 9.1855 | 3.7725 | 2.0645 | 1.3765 |
| 2000 | 147.8075 | 41.0165 | 5.4855 | 2.2645 | 1.0855 | 0.9295 | 0.7035 |
| 2001 | 45.5375 | 31.4445 | 19.1445 | 4.7515 | 3.5365 | 1.2165 | 1.6635 |
| 2002 | 127.0935 | 53.1605 | 14.2155 | 3.2315 | 0.4025 | 1.0425 | 0.0895 |
| 2003 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 2004 | 186.5145 | 27.0285 | 18.7565 | 4.0895 | 2.9985 | 3.4225 | 0.0895 |
| 2005 | 75.3905 | 155.5475 | 23.6635 | 0.0895 | 0.0895 | 0.0895 | 0.0895 |
| 2006 | 196.1535 | 97.4725 | 14.8685 | 3.6135 | 1.0895 | 0.0895 | 0.0895 |
| 2007 | 89.7415 | 55.6055 | 33.7815 | 11.8455 | 1.3245 | 0.0895 | 0.0895 |
| 2008 | 52.0905 | 99.7425 | 40.8285 | 11.8675 | 10.9225 | 1.2005 | 7.4825 |
| 2009 | 26.2675 | 20.3115 | 5.6455 | 14.4675 | 5.0895 | 0.0895 | 0.0895 |
| 2010 | 96.0185 | 35.8115 | 9.2565 | 5.3675 | 3.7005 | 6.7565 | 1.2005 |
| 2011 | 116.6895 | 36.8895 | 0.0895 | 0.0895 | 0.0895 | 1.6895 | 0.0895 |
| 2012 | 39.8585 | 33.5115 | 9.4645 | 1.2325 | 0.0895 | 0.0895 | 0.0895 |
| 2013 | 110.1595 | 16.1155 | 15.6395 | 0.4405 | 0.0895 | 0.0895 | 0.0895 |
| 2014 | 102.7143 | 18.3059 | 9.4471 | 6.1647 | 4.7412 | 1.2 | 0.9412 |
| 2015 | 273.794 | 45.873 | 2.000 | 2.000 | 0.0895 | 0.0895 | 0.0895 |

Table 18.2.2 cont. Turbot in Subarea 4: Relative abundance indices used in the assessment. BTS-ISIS.

| BTS-ISIS |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Start | 0.66 | End | 0.75 |  |  |  |  |
|  | 1 | 2 | 3 | 4 | 5 |  |  |
| 1985 | 0.39545 | 1.28365 | 0.26525 | 0.11355 | 0.05375 | 0.02135 | 0.00955 |
| 1986 | 0.22755 | 0.90425 | 0.27645 | 0.10075 | 0.04255 | 0.01925 | 0.00745 |
| 1987 | 0.26385 | 1.08715 | 0.27415 | 0.11135 | 0.05005 | 0.03425 | 0.00765 |
| 1988 | 0.58935 | 1.19375 | 0.30315 | 0.08525 | 0.03035 | 0.02755 | 0.00745 |
| 1989 | 0.37815 | 1.35885 | 0.39915 | 0.12165 | 0.03985 | 0.02615 | 0.01545 |
| 1990 | 2.08255 | 1.24465 | 0.25975 | 0.12565 | 0.06665 | 0.02745 | 0.00965 |
| 1991 | 1.22675 | 1.66485 | 0.21705 | 0.02365 | 0.01415 | 0.00025 | 0.01215 |
| 1992 | 1.36115 | 1.17845 | 0.31985 | 0.03375 | 0.01545 | 0.01055 | 0.00335 |
| 1993 | 1.67955 | 1.40555 | 0.18545 | 0.05225 | 0.04505 | 0.00175 | 0.00075 |
| 1994 | 1.83025 | 1.58005 | 0.10225 | 0.03125 | 0.00625 | 0.00325 | 0.00325 |
| 1995 | 1.83265 | 0.60705 | 0.10125 | 0.01185 | 0.00895 | 0.00345 | 0.00025 |
| 1996 | 0.61465 | 1.90125 | 0.11255 | 0.07465 | 0.04045 | 0.00025 | 0.00915 |
| 1997 | 0.66885 | 1.30755 | 0.37765 | 0.02625 | 0.03805 | 0.01335 | 0.01155 |
| 1998 | 1.91495 | 0.91595 | 0.23285 | 0.15245 | 0.00475 | 0.00025 | 0.00135 |
| 1999 | 1.24255 | 1.18095 | 0.19545 | 0.09545 | 0.01665 | 0.00265 | 0.00115 |
| 2000 | 4.21375 | 0.84715 | 0.38565 | 0.16375 | 0.05395 | 0.05465 | 0.00025 |
| 2001 | 1.04385 | 1.40955 | 0.12875 | 0.15225 | 0.00025 | 0.00025 | 0.04025 |
| 2002 | 2.81445 | 0.49315 | 0.14595 | 0.04625 | 0.03195 | 0.02175 | 0.00095 |
| 2003 | 1.54345 | 0.87475 | 0.10115 | 0.05435 | 0.00025 | 0.01215 | 0.01145 |
| 2004 | 2.16585 | 0.63995 | 0.35895 | 0.00025 | 0.06855 | 0.01715 | 0.00025 |
| 2005 | 1.14255 | 1.53825 | 0.52595 | 0.11575 | 0.03595 | 0.00625 | 0.01215 |
| 2006 | 1.70525 | 0.79935 | 0.27315 | 0.11375 | 0.00475 | 0.00025 | 0.00025 |
| 2007 | 1.34235 | 0.90235 | 0.56285 | 0.27955 | 0.09035 | 0.06005 | 0.00025 |
| 2008 | 1.19555 | 1.12475 | 0.43125 | 0.14325 | 0.07615 | 0.01735 | 0.07975 |
| 2009 | 0.97165 | 0.41985 | 0.34585 | 0.28145 | 0.15225 | 0.04955 | 0.00505 |
| 2010 | 1.69095 | 0.34825 | 0.09925 | 0.07015 | 0.08895 | 0.01465 | 0.01465 |
| 2011 | 1.84005 | 0.89155 | 0.16335 | 0.06325 | 0.06535 | 0.01665 | 0.00025 |
| 2012 | 0.97725 | 0.93035 | 0.24015 | 0.23555 | 0.02135 | 0.04495 | 0.08375 |
| 2013 | 0.66795 | 0.58505 | 0.45565 | 0.15835 | 0.01775 | 0.03735 | 0.04055 |
| 2014 | 1.179 | 0.173 | 0.070 | 0.087 | 0.00025 | 0.00025 |  |
| 2015 |  |  |  |  |  |  |  |

Table 18.2.2 cont. Turbot in Subarea 4: Relative abundance indices used in the assessment. Dutch 80 mm beam trawl (BT2) LPUE: age-aggregated index of exploitable biomass per unit effort (top) and the age-disaggregated index of abundance per unit effort.

|  |  | DutCH_BT2_LPUE |
| :--- | :--- | :--- |
| start | 0 |  |
| end | 1 |  |
|  | Exploitable biomass |  |
| 2002 | 66.545 |  |
| 2003 | 68.835 |  |
| 2004 | 70.225 |  |
| 2005 | 67.795 |  |
| 2006 | 69.505 |  |
| 2007 | 89.185 |  |
| 2008 | 102.285 |  |
| 2009 | 105.585 |  |
| 2010 | 86.595 |  |
| 2011 | 97.285 |  |
| 2012 | 93.515 |  |
| 2013 | 105.955 |  |
| 2014 | 89.775 |  |
| 2015 | 94.895 |  |

DUTCH_BT2_LPUE_AGES

| Start | 0 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| End | 1 |  |  |  |  |  |  |  |  |
|  | 1 | 2 | 3 | 4 | 8 | 9 |  |  |  |
| 2002 | 1.833 | 13.993 | 30.153 | 8.703 | 6.053 | 2.083 | 0.753 | 0.963 | 0.173 |
| 2003 | 2.383 | 25.533 | 14.613 | 15.043 | 4.173 | 2.643 | 1.573 | 0.533 | 0.723 |
| 2004 | 3.103 | 27.583 | 23.873 | 5.273 | 5.903 | 1.223 | 0.953 | 0.333 | 0.153 |
| 2005 | 1.413 | 23.213 | 25.313 | 10.863 | 1.793 | 2.793 | 0.173 | 0.483 | 0.193 |
| 2006 | 4.193 | 20.453 | 27.443 | 8.333 | 4.073 | 1.263 | 1.583 | 0.403 | 0.053 |
| 2007 | 0.993 | 42.923 | 20.493 | 13.743 | 3.793 | 2.753 | 0.483 | 0.703 | 0.023 |
| 2008 | 2.733 | 30.313 | 34.123 | 11.163 | 11.983 | 3.563 | 2.523 | 0.283 | 0.743 |
| 2009 | 0.903 | 17.133 | 37.753 | 28.783 | 8.503 | 4.113 | 1.953 | 1.313 | 0.153 |
| 2010 | 2.793 | 22.883 | 14.933 | 15.733 | 11.893 | 5.613 | 2.623 | 1.133 | 0.903 |
| 2011 | 4.523 | 29.833 | 25.053 | 8.823 | 10.633 | 8.713 | 3.473 | 1.393 | 0.563 |
| 2012 | 4.343 | 35.973 | 22.713 | 11.723 | 3.823 | 5.573 | 5.073 | 0.893 | 0.913 |
| 2013 | 3.693 | 24.553 | 39.023 | 17.583 | 7.613 | 4.623 | 6.373 | 3.813 | 0.633 |
| 2014 | 1.855 | 14.615 | 26.725 | 23.185 | 10.465 | 6.155 | 2.025 | 2.575 | 2.175 |
| 2015 | 0.875 | 33.225 | 17.015 | 12.325 | 19.635 | 6.835 | 2.055 | 1.295 | 1.635 |

Table 18.2.3 Turbot in Subarea 4: Biological data used in the assessment. Maturity and natural mortality values are constant over years.

|  |  | AGE |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| Maturit |  | 0 | 0.04 | 0.47 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| Natural Mor |  | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| Sтоск W@A | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |  | 9 | 10 |
| 1975 | 0.27 | 0.58 | 1.01 | 1.53 | 2.12 | 2.75 | 3.41 | 4.08 |  | 4.75 | 5.39 |
| 1976 | 0.28 | 0.60 | 1.04 | 1.57 | 2.18 | 2.84 | 3.52 | 4.21 |  | 4.90 | 5.56 |
| 1977 | 0.29 | 0.62 | 1.07 | 1.63 | 2.25 | 2.93 | 3.64 | 4.35 |  | 5.06 | 5.75 |
| 1978 | 0.30 | 0.64 | 1.11 | 1.68 | 2.33 | 3.03 | 3.76 | 4.50 |  | 5.23 | 5.94 |
| 1979 | 0.31 | 0.67 | 1.15 | 1.74 | 2.41 | 3.14 | 3.89 | 4.66 |  | 5.42 | 6.15 |
| 1980 | 0.32 | 0.69 | 1.19 | 1.80 | 2.50 | 3.25 | 4.03 | 4.82 |  | 5.61 | 6.37 |
| 1981 | 0.33 | 0.71 | 1.23 | 1.87 | 2.59 | 3.37 | 4.18 | 5.00 |  | 5.81 | 6.60 |
| 1982 | 0.34 | 0.74 | 1.28 | 1.93 | 2.68 | 3.48 | 4.32 | 5.17 |  | 6.01 | 6.83 |
| 1983 | 0.36 | 0.76 | 1.32 | 2.00 | 2.77 | 3.60 | 4.47 | 5.35 |  | 6.22 | 7.06 |
| 1984 | 0.37 | 0.79 | 1.36 | 2.07 | 2.86 | 3.72 | 4.62 | 5.53 |  | 6.42 | 7.30 |
| 1985 | 0.38 | 0.81 | 1.41 | 2.13 | 2.95 | 3.84 | 4.76 | 5.70 |  | 6.63 | 7.53 |
| 1986 | 0.39 | 0.84 | 1.45 | 2.19 | 3.04 | 3.95 | 4.90 | 5.87 |  | 6.82 | 7.75 |
| 1987 | 0.40 | 0.86 | 1.49 | 2.25 | 3.12 | 4.06 | 5.04 | 6.02 |  | 7.00 | 7.96 |
| 1988 | 0.41 | 0.88 | 1.52 | 2.31 | 3.20 | 4.16 | 5.16 | 6.17 |  | 7.17 | 8.15 |
| 1989 | 0.42 | 0.90 | 1.55 | 2.35 | 3.26 | 4.24 | 5.26 | 6.30 |  | 7.32 | 8.32 |
| 1990 | 0.43 | 0.91 | 1.58 | 2.39 | 3.32 | 4.32 | 5.35 | 6.41 |  | 7.45 | 8.46 |
| 1991 | 0.43 | 0.93 | 1.60 | 2.43 | 3.36 | 4.37 | 5.42 | 6.49 |  | 7.54 | 8.57 |
| 1992 | 0.44 | 0.93 | 1.62 | 2.45 | 3.39 | 4.41 | 5.47 | 6.55 |  | 7.61 | 8.64 |
| 1993 | 0.44 | 0.94 | 1.62 | 2.46 | 3.40 | 4.43 | 5.49 | 6.57 |  | 7.64 | 8.68 |
| 1994 | 0.44 | 0.94 | 1.62 | 2.45 | 3.40 | 4.42 | 5.48 | 6.56 |  | 7.63 | 8.67 |
| 1995 | 0.43 | 0.93 | 1.61 | 2.44 | 3.37 | 4.39 | 5.45 | 6.52 |  | 7.58 | 8.61 |
| 1996 | 0.43 | 0.92 | 1.59 | 2.40 | 3.33 | 4.33 | 5.38 | 6.43 |  | 7.48 | 8.50 |
| 1997 | 0.42 | 0.90 | 1.56 | 2.36 | 3.27 | 4.26 | 5.28 | 6.32 |  | 7.34 | 8.34 |
| 1998 | 0.41 | 0.88 | 1.52 | 2.31 | 3.20 | 4.16 | 5.16 | 6.17 |  | 7.18 | 8.15 |
| 1999 | 0.40 | 0.86 | 1.48 | 2.25 | 3.11 | 4.05 | 5.02 | 6.01 |  | 6.99 | 7.94 |
| 2000 | 0.39 | 0.83 | 1.44 | 2.18 | 3.02 | 3.93 | 4.87 | 5.83 |  | 6.78 | 7.70 |
| 2001 | 0.38 | 0.80 | 1.39 | 2.11 | 2.92 | 3.80 | 4.71 | 5.64 |  | 6.55 | 7.44 |
| 2002 | 0.36 | 0.78 | 1.34 | 2.03 | 2.82 | 3.66 | 4.54 | 5.44 |  | 6.32 | 7.18 |
| 2003 | 0.35 | 0.75 | 1.29 | 1.96 | 2.71 | 3.53 | 4.38 | 5.23 |  | 6.09 | 6.91 |
| 2004 | 0.34 | 0.72 | 1.24 | 1.88 | 2.61 | 3.39 | 4.21 | 5.04 |  | 5.85 | 6.65 |
| 2005 | 0.32 | 0.69 | 1.19 | 1.81 | 2.51 | 3.26 | 4.05 | 4.84 |  | 5.63 | 6.39 |
| 2006 | 0.31 | 0.66 | 1.15 | 1.74 | 2.41 | 3.14 | 3.89 | 4.65 |  | 5.41 | 6.15 |
| 2007 | 0.30 | 0.64 | 1.11 | 1.67 | 2.32 | 3.02 | 3.74 | 4.48 |  | 5.21 | 5.92 |
| 2008 | 0.29 | 0.62 | 1.06 | 1.61 | 2.23 | 2.91 | 3.61 | 4.31 |  | 5.02 | 5.70 |
| 2009 | 0.28 | 0.59 | 1.03 | 1.56 | 2.16 | 2.81 | 3.48 | 4.16 |  | 4.84 | 5.50 |
| 2010 | 0.27 | 0.58 | 0.99 | 1.51 | 2.09 | 2.72 | 3.37 | 4.03 |  | 4.69 | 5.32 |
| 2011 | 0.26 | 0.56 | 0.97 | 1.46 | 2.03 | 2.64 | 3.27 | 3.91 |  | 4.55 | 5.17 |
| 2012 | 0.25 | 0.54 | 0.94 | 1.42 | 1.97 | 2.57 | 3.19 | 3.81 |  | 4.43 | 5.03 |
| 2013 | 0.25 | 0.53 | 0.92 | 1.39 | 1.93 | 2.51 | 3.12 | 3.73 |  | 4.34 | 4.93 |


| 2014 | 0.24 | 0.52 | 0.91 | 1.37 | 1.90 | 2.47 | 3.07 | 3.67 | 4.26 | 4.84 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 0.24 | 0.52 | 0.90 | 1.36 | 1.88 | 2.44 | 3.03 | 3.63 | 4.22 | 4.79 |
| Landings W@A | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1975 | 0.28 | 0.61 | 1.05 | 1.59 | 2.20 | 2.86 | 3.55 | 4.25 | 4.94 | 5.61 |
| 1976 | 0.29 | 0.63 | 1.08 | 1.64 | 2.27 | 2.95 | 3.66 | 4.38 | 5.09 | 5.79 |
| 1977 | 0.30 | 0.65 | 1.12 | 1.69 | 2.34 | 3.05 | 3.78 | 4.53 | 5.26 | 5.98 |
| 1978 | 0.31 | 0.67 | 1.16 | 1.75 | 2.42 | 3.15 | 3.91 | 4.68 | 5.44 | 6.18 |
| 1979 | 0.32 | 0.69 | 1.20 | 1.81 | 2.51 | 3.26 | 4.05 | 4.84 | 5.63 | 6.40 |
| 1980 | 0.33 | 0.72 | 1.24 | 1.88 | 2.60 | 3.38 | 4.19 | 5.02 | 5.83 | 6.63 |
| 1981 | 0.35 | 0.74 | 1.28 | 1.94 | 2.69 | 3.50 | 4.34 | 5.20 | 6.04 | 6.86 |
| 1982 | 0.36 | 0.77 | 1.33 | 2.01 | 2.79 | 3.62 | 4.50 | 5.38 | 6.25 | 7.10 |
| 1983 | 0.37 | 0.79 | 1.37 | 2.08 | 2.88 | 3.75 | 4.65 | 5.56 | 6.47 | 7.35 |
| 1984 | 0.38 | 0.82 | 1.42 | 2.15 | 2.98 | 3.87 | 4.80 | 5.75 | 6.68 | 7.59 |
| 1985 | 0.40 | 0.85 | 1.46 | 2.22 | 3.07 | 3.99 | 4.95 | 5.93 | 6.89 | 7.83 |
| 1986 | 0.41 | 0.87 | 1.51 | 2.28 | 3.16 | 4.11 | 5.10 | 6.10 | 7.09 | 8.06 |
| 1987 | 0.42 | 0.89 | 1.55 | 2.34 | 3.25 | 4.22 | 5.24 | 6.27 | 7.28 | 8.28 |
| 1988 | 0.43 | 0.92 | 1.58 | 2.40 | 3.32 | 4.32 | 5.36 | 6.42 | 7.46 | 8.47 |
| 1989 | 0.44 | 0.94 | 1.62 | 2.45 | 3.39 | 4.41 | 5.47 | 6.55 | 7.61 | 8.65 |
| 1990 | 0.44 | 0.95 | 1.64 | 2.49 | 3.45 | 4.49 | 5.57 | 6.66 | 7.74 | 8.80 |
| 1991 | 0.45 | 0.96 | 1.67 | 2.52 | 3.50 | 4.55 | 5.64 | 6.75 | 7.85 | 8.91 |
| 1992 | 0.45 | 0.97 | 1.68 | 2.54 | 3.53 | 4.59 | 5.69 | 6.81 | 7.91 | 8.99 |
| 1993 | 0.46 | 0.98 | 1.69 | 2.55 | 3.54 | 4.60 | 5.71 | 6.83 | 7.94 | 9.03 |
| 1994 | 0.46 | 0.97 | 1.68 | 2.55 | 3.53 | 4.60 | 5.70 | 6.82 | 7.93 | 9.01 |
| 1995 | 0.45 | 0.97 | 1.67 | 2.53 | 3.51 | 4.57 | 5.66 | 6.78 | 7.88 | 8.95 |
| 1996 | 0.45 | 0.96 | 1.65 | 2.50 | 3.47 | 4.51 | 5.59 | 6.69 | 7.78 | 8.84 |
| 1997 | 0.44 | 0.94 | 1.62 | 2.46 | 3.40 | 4.43 | 5.49 | 6.57 | 7.64 | 8.68 |
| 1998 | 0.43 | 0.92 | 1.58 | 2.40 | 3.33 | 4.33 | 5.37 | 6.42 | 7.47 | 8.48 |
| 1999 | 0.42 | 0.89 | 1.54 | 2.34 | 3.24 | 4.21 | 5.22 | 6.25 | 7.27 | 8.25 |
| 2000 | 0.40 | 0.87 | 1.50 | 2.27 | 3.14 | 4.08 | 5.07 | 6.06 | 7.05 | 8.01 |
| 2001 | 0.39 | 0.84 | 1.45 | 2.19 | 3.04 | 3.95 | 4.90 | 5.86 | 6.81 | 7.74 |
| 2002 | 0.38 | 0.81 | 1.40 | 2.11 | 2.93 | 3.81 | 4.73 | 5.65 | 6.57 | 7.47 |
| 2003 | 0.36 | 0.78 | 1.34 | 2.03 | 2.82 | 3.67 | 4.55 | 5.44 | 6.33 | 7.19 |
| 2004 | 0.35 | 0.75 | 1.29 | 1.96 | 2.71 | 3.53 | 4.38 | 5.24 | 6.09 | 6.92 |
| 2005 | 0.34 | 0.72 | 1.24 | 1.88 | 2.61 | 3.39 | 4.21 | 5.03 | 5.85 | 6.65 |
| 2006 | 0.32 | 0.69 | 1.19 | 1.81 | 2.51 | 3.26 | 4.05 | 4.84 | 5.63 | 6.39 |
| 2007 | 0.31 | 0.67 | 1.15 | 1.74 | 2.41 | 3.14 | 3.89 | 4.66 | 5.41 | 6.15 |
| 2008 | 0.30 | 0.64 | 1.11 | 1.68 | 2.32 | 3.02 | 3.75 | 4.49 | 5.22 | 5.93 |
| 2009 | 0.29 | 0.62 | 1.07 | 1.62 | 2.24 | 2.92 | 3.62 | 4.33 | 5.04 | 5.72 |
| 2010 | 0.28 | 0.60 | 1.03 | 1.57 | 2.17 | 2.82 | 3.50 | 4.19 | 4.87 | 5.54 |
| 2011 | 0.27 | 0.58 | 1.00 | 1.52 | 2.11 | 2.74 | 3.40 | 4.07 | 4.73 | 5.37 |
| 2012 | 0.26 | 0.57 | 0.98 | 1.48 | 2.05 | 2.67 | 3.31 | 3.96 | 4.61 | 5.24 |
| 2013 | 0.26 | 0.55 | 0.96 | 1.45 | 2.01 | 2.61 | 3.24 | 3.88 | 4.51 | 5.12 |
| 2014 | 0.25 | 0.54 | 0.94 | 1.43 | 1.98 | 2.57 | 3.19 | 3.81 | 4.43 | 5.04 |
| 2015 | 0.25 | 0.54 | 0.93 | 1.41 | 1.95 | 2.54 | 3.15 | 3.77 | 4.38 | 4.98 |

Table 18.4.1 Turbot in Subarea 4: Estimates of stock numbers at age (thousands) from the SAM assessment.

| N@A | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 4744 | 4913 | 2827 | 881 | 545 | 540 | 430 | 261 | 231 | 918 |
| 1976 | 6898 | 3766 | 3573 | 1466 | 501 | 341 | 338 | 278 | 170 | 749 |
| 1977 | 8518 | 5899 | 2665 | 1885 | 835 | 309 | 218 | 221 | 182 | 598 |
| 1978 | 4683 | 7212 | 4269 | 1472 | 1103 | 530 | 207 | 148 | 147 | 513 |
| 1979 | 3195 | 3892 | 5298 | 2346 | 882 | 679 | 354 | 138 | 99 | 438 |
| 1980 | 3815 | 2594 | 2789 | 2896 | 1347 | 557 | 427 | 238 | 91 | 354 |
| 1981 | 3453 | 3153 | 1808 | 1422 | 1629 | 823 | 359 | 273 | 159 | 291 |
| 1982 | 5017 | 2801 | 2256 | 836 | 724 | 971 | 514 | 236 | 180 | 298 |
| 1983 | 6274 | 4173 | 2027 | 1072 | 411 | 418 | 587 | 333 | 149 | 306 |
| 1984 | 5207 | 5286 | 2916 | 991 | 508 | 223 | 241 | 354 | 204 | 279 |
| 1985 | 2888 | 4424 | 3709 | 1303 | 500 | 266 | 124 | 148 | 218 | 298 |
| 1986 | 3050 | 2310 | 3040 | 1399 | 530 | 259 | 132 | 74 | 88 | 311 |
| 1987 | 3974 | 2608 | 1525 | 1281 | 605 | 269 | 154 | 76 | 45 | 250 |
| 1988 | 4523 | 3417 | 1826 | 710 | 659 | 355 | 168 | 112 | 52 | 199 |
| 1989 | 4756 | 3711 | 2249 | 837 | 382 | 385 | 217 | 109 | 72 | 162 |
| 1990 | 6091 | 4017 | 2446 | 976 | 428 | 217 | 248 | 147 | 70 | 151 |
| 1991 | 5664 | 4957 | 2576 | 988 | 455 | 214 | 120 | 144 | 85 | 128 |
| 1992 | 6289 | 4715 | 3066 | 984 | 453 | 229 | 118 | 70 | 84 | 125 |
| 1993 | 5761 | 5144 | 2927 | 1027 | 410 | 219 | 123 | 68 | 41 | 121 |
| 1994 | 4852 | 4685 | 2977 | 923 | 386 | 183 | 115 | 70 | 39 | 92 |
| 1995 | 5915 | 3807 | 2668 | 814 | 336 | 163 | 93 | 65 | 40 | 74 |
| 1996 | 4270 | 4939 | 2243 | 832 | 316 | 160 | 86 | 55 | 38 | 67 |
| 1997 | 3430 | 3516 | 2945 | 876 | 385 | 166 | 98 | 52 | 35 | 67 |
| 1998 | 4401 | 2774 | 2165 | 1224 | 445 | 217 | 103 | 65 | 33 | 67 |
| 1999 | 3967 | 3710 | 1670 | 943 | 626 | 276 | 144 | 72 | 45 | 69 |
| 2000 | 4697 | 3094 | 2159 | 632 | 419 | 316 | 167 | 87 | 44 | 70 |
| 2001 | 2979 | 3844 | 1686 | 691 | 217 | 145 | 127 | 70 | 35 | 47 |
| 2002 | 4345 | 2285 | 1988 | 424 | 225 | 82 | 51 | 44 | 24 | 29 |
| 2003 | 3878 | 3519 | 1031 | 605 | 133 | 90 | 39 | 22 | 19 | 23 |
| 2004 | 5608 | 2985 | 1477 | 218 | 179 | 42 | 38 | 17 | 10 | 18 |
| 2005 | 4855 | 4591 | 1195 | 378 | 68 | 79 | 20 | 22 | 9 | 16 |
| 2006 | 6069 | 3842 | 1947 | 383 | 156 | 33 | 47 | 11 | 13 | 15 |
| 2007 | 5215 | 4965 | 1803 | 835 | 190 | 94 | 20 | 30 | 7 | 17 |
| 2008 | 2974 | 4376 | 2401 | 771 | 400 | 105 | 56 | 11 | 18 | 14 |
| 2009 | 3362 | 2254 | 2011 | 1080 | 370 | 191 | 54 | 32 | 6 | 19 |
| 2010 | 4951 | 2712 | 919 | 807 | 522 | 200 | 107 | 31 | 19 | 15 |
| 2011 | 5339 | 4060 | 1206 | 364 | 382 | 283 | 104 | 60 | 17 | 19 |
| 2012 | 3591 | 4419 | 1754 | 523 | 171 | 201 | 155 | 56 | 33 | 20 |
| 2013 | 2699 | 2950 | 2027 | 740 | 246 | 96 | 110 | 83 | 31 | 30 |
| 2014 | 5061 | 1991 | 1350 | 853 | 332 | 134 | 57 | 62 | 48 | 36 |
| 2015 | 8038 | 4073 | 833 | 504 | 293 | 135 | 58 | 29 | 30 | 41 |

Table 18.4.2. Turbot in Subarea 4: Estimates of fishing mortality at age from the SAM assessment.

| F@A | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 0.000 | 0.131 | 0.449 | 0.363 | 0.279 | 0.255 | 0.234 |
| 1976 | 0.000 | 0.132 | 0.444 | 0.363 | 0.280 | 0.248 | 0.230 |
| 1977 | 0.000 | 0.126 | 0.396 | 0.327 | 0.252 | 0.217 | 0.206 |
| 1978 | 0.000 | 0.127 | 0.398 | 0.328 | 0.256 | 0.218 | 0.205 |
| 1979 | 0.000 | 0.130 | 0.421 | 0.354 | 0.271 | 0.232 | 0.208 |
| 1980 | 0.000 | 0.138 | 0.469 | 0.397 | 0.296 | 0.252 | 0.215 |
| 1981 | 0.000 | 0.147 | 0.532 | 0.458 | 0.329 | 0.275 | 0.221 |
| 1982 | 0.000 | 0.150 | 0.543 | 0.488 | 0.360 | 0.318 | 0.256 |
| 1983 | 0.001 | 0.157 | 0.553 | 0.520 | 0.399 | 0.362 | 0.297 |
| 1984 | 0.001 | 0.173 | 0.600 | 0.536 | 0.414 | 0.368 | 0.288 |
| 1985 | 0.001 | 0.207 | 0.766 | 0.682 | 0.503 | 0.431 | 0.317 |
| 1986 | 0.001 | 0.208 | 0.705 | 0.624 | 0.456 | 0.363 | 0.270 |
| 1987 | 0.001 | 0.197 | 0.579 | 0.477 | 0.341 | 0.257 | 0.185 |
| 1988 | 0.001 | 0.210 | 0.597 | 0.467 | 0.364 | 0.292 | 0.233 |
| 1989 | 0.001 | 0.219 | 0.616 | 0.474 | 0.375 | 0.290 | 0.235 |
| 1990 | 0.001 | 0.248 | 0.718 | 0.569 | 0.484 | 0.401 | 0.344 |
| 1991 | 0.001 | 0.265 | 0.765 | 0.595 | 0.492 | 0.398 | 0.336 |
| 1992 | 0.002 | 0.295 | 0.873 | 0.673 | 0.537 | 0.420 | 0.347 |
| 1993 | 0.002 | 0.323 | 0.968 | 0.755 | 0.593 | 0.449 | 0.362 |
| 1994 | 0.002 | 0.355 | 1.078 | 0.823 | 0.637 | 0.471 | 0.372 |
| 1995 | 0.002 | 0.346 | 0.971 | 0.740 | 0.561 | 0.415 | 0.331 |
| 1996 | 0.002 | 0.309 | 0.759 | 0.575 | 0.434 | 0.320 | 0.262 |
| 1997 | 0.002 | 0.297 | 0.676 | 0.502 | 0.370 | 0.270 | 0.222 |
| 1998 | 0.002 | 0.293 | 0.623 | 0.459 | 0.328 | 0.234 | 0.192 |
| 1999 | 0.003 | 0.338 | 0.748 | 0.597 | 0.474 | 0.361 | 0.324 |
| 2000 | 0.004 | 0.409 | 0.969 | 0.865 | 0.804 | 0.697 | 0.700 |
| 2001 | 0.005 | 0.484 | 1.128 | 0.951 | 0.840 | 0.788 | 0.845 |
| 2002 | 0.006 | 0.526 | 1.041 | 0.892 | 0.733 | 0.595 | 0.606 |
| 2003 | 0.009 | 0.659 | 1.317 | 1.083 | 0.895 | 0.672 | 0.639 |
| 2004 | 0.009 | 0.671 | 1.158 | 0.917 | 0.643 | 0.450 | 0.388 |
| 2005 | 0.009 | 0.607 | 0.890 | 0.686 | 0.479 | 0.360 | 0.331 |
| 2006 | 0.007 | 0.528 | 0.626 | 0.479 | 0.352 | 0.306 | 0.305 |
| 2007 | 0.008 | 0.557 | 0.632 | 0.505 | 0.396 | 0.361 | 0.347 |
| 2008 | 0.009 | 0.583 | 0.639 | 0.526 | 0.438 | 0.388 | 0.334 |
| 2009 | 0.010 | 0.651 | 0.705 | 0.540 | 0.411 | 0.363 | 0.330 |
| 2010 | 0.010 | 0.651 | 0.699 | 0.552 | 0.426 | 0.398 | 0.359 |
| 2011 | 0.010 | 0.648 | 0.694 | 0.574 | 0.444 | 0.420 | 0.392 |
| 2012 | 0.010 | 0.622 | 0.665 | 0.578 | 0.432 | 0.414 | 0.388 |
| 2013 | 0.010 | 0.625 | 0.674 | 0.612 | 0.447 | 0.402 | 0.353 |
| 2014 | 0.011 | 0.693 | 0.831 | 0.856 | 0.674 | 0.631 | 0.523 |
| 2015 | 0.013 | 0.772 | 1.077 | 1.184 | 0.966 | 0.929 | 0.761 |

Table 18.4.3. Turbot in Subarea 4: Summary of assessment results. Values by year with the geometric mean over all years. Biomass values are in tons ( t ).

|  | Rec | TSB | SSB | Official Land | Mean F (ages 2-6) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 4744 | 19573 | 13952 | 4589 | 0.295 |
| 1976 | 6898 | 19649 | 13452 | 4816 | 0.293 |
| 1977 | 8518 | 20967 | 13304 | 4486 | 0.263 |
| 1978 | 4683 | 22697 | 14204 | 5036 | 0.265 |
| 1979 | 3195 | 23272 | 16352 | 6365 | 0.282 |
| 1980 | 3815 | 22382 | 17407 | 5486 | 0.31 |
| 1981 | 3453 | 20986 | 16360 | 4756 | 0.348 |
| 1982 | 5017 | 20177 | 14854 | 4454 | 0.372 |
| 1983 | 6274 | 20384 | 13559 | 4576 | 0.398 |
| 1984 | 5207 | 20810 | 12676 | 5297 | 0.418 |
| 1985 | 2888 | 20309 | 12851 | 6188 | 0.518 |
| 1986 | 3050 | 17320 | 11782 | 5264 | 0.471 |
| 1987 | 3974 | 15511 | 10413 | 4272 | 0.37 |
| 1988 | 4523 | 16411 | 10104 | 4042 | 0.386 |
| 1989 | 4756 | 17382 | 10228 | 4927 | 0.395 |
| 1990 | 6091 | 18907 | 10611 | 5751 | 0.484 |
| 1991 | 5664 | 19357 | 10189 | 6340 | 0.503 |
| 1992 | 6289 | 19887 | 10164 | 5934 | 0.56 |
| 1993 | 5761 | 19467 | 9667 | 5547 | 0.617 |
| 1994 | 4852 | 17912 | 8903 | 5244 | 0.673 |
| 1995 | 5915 | 16102 | 7757 | 4672 | 0.607 |
| 1996 | 4270 | 15349 | 7172 | 3644 | 0.479 |
| 1997 | 3430 | 14904 | 7875 | 3382 | 0.423 |
| 1998 | 4401 | 14426 | 8375 | 3087 | 0.387 |
| 1999 | 3967 | 14452 | 8387 | 3187 | 0.503 |
| 2000 | 4697 | 13549 | 7536 | 4026 | 0.749 |
| 2001 | 2979 | 10771 | 5366 | 4101 | 0.838 |
| 2002 | 4345 | 8634 | 3899 | 3750 | 0.757 |
| 2003 | 3878 | 7744 | 3100 | 3375 | 0.925 |
| 2004 | 5608 | 7303 | 2365 | 3319 | 0.768 |
| 2005 | 4855 | 7620 | 2213 | 3195 | 0.604 |
| 2006 | 6069 | 8220 | 2665 | 2977 | 0.458 |
| 2007 | 5215 | 9193 | 3460 | 3510 | 0.49 |
| 2008 | 2974 | 8966 | 4104 | 3007 | 0.515 |
| 2009 | 3362 | 7818 | 4417 | 3091 | 0.534 |
| 2010 | 4951 | 7309 | 3933 | 2692 | 0.545 |
| 2011 | 5339 | 7633 | 3419 | 2807 | 0.556 |
| 2012 | 3591 | 7521 | 3388 | 2914 | 0.542 |
| 2013 | 2699 | 6790 | 3570 | 3084 | 0.552 |
| 2014 | 5061 | 6414 | 3469 | 2834 | 0.737 |
| 2015 | 8038 | 6967 | 2569 | 2925 | 0.986 |
| Geo.Mean | 4586 | 13451 | 7177 |  |  |

Table 18.5.1. Standard deviation estimates for random walks and observations.

| Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Random walk $\log (\mathrm{F})$ SDs |  |  |  |  |  |  |  |  |  |  |
|  | 0.29 | 0.18 | 0.29 | 0.29 | 0.29 | 0.29 | 0.29 | 0.29 | 0.29 | 0.29 |
| Random walk $\log (\mathrm{N})$ SDs |  |  |  |  |  |  |  |  |  |  |
|  | 0.37 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 |
| Rho correlation parameter for random walks on F at age |  |  |  |  |  |  |  |  |  |  |
| 0.86 |  |  |  |  |  |  |  |  |  |  |
| Observation SDs |  |  |  |  |  |  |  |  |  |  |
| Landings | 2.19 | 0.55 | 0.31 | 0.31 | 0.31 | 0.31 | 0.31 | 0.31 | 0.61 | 0.61 |
| SNS | 0.64 | 0.50 | 1.10 | 1.10 | 1.10 | 1.10 | 1.10 |  |  |  |
| BTS | 0.62 | 0.32 | 1.43 | 1.43 | 1.43 | 1.43 | 1.43 |  |  |  |



Figure 18.2.1. Turbot in Subarea 4. Top: Total landings by metier in 2014 sorted by sampled/unsampled for numbers at age in InterCatch. Bottom: Distributions of numbers at age by quarter for the two metiers with available samples - NL TBB_DEF_70-99 (left) and NL OTB_DEF_70-99 (right).


Figure 18.2.2. Turbot in Subarea 4. Total landings 1957-2014 (from the ICES database of official landings).


Figure 18.2.3. Turbot in Subarea 4. Official landings by country in 2014 (from the ICES database of official landings) for turbot (blue), turbot and brill combined (green) compared to the combined TAC for turbot and brill (orange).


Figure 18.2.4. Turbot in Subarea 4. Landings at age for the years with available data between 19852014.


Figure 18.2.5a. Turbot in Subarea 4. Time series of the standardized indices for ages 1 to 9 from the three tuning fleets available for the assessment: BTS-ISIS (black), SNS (red) and NL beam trawl LPUE (blue; not used in the final assessment).


Figure 18.2.5b. Turbot in Subarea 4. Time series of exploitable biomass per unit effort from the NL 80 mm beam trawl fleet.


Figure 18.2.6 Turbot in Subarea 4. Internal consistency of the three tuning indices available for the assessment : BTS-ISIS (top), SNS (bottom) and NL beam trawl LPUE (next page, not used in the final assessment).

$\log _{10}$ (Index Value)
Low er right panels show the Coefficient of Determination $\left(r^{2}\right)$

Figure 18.2.6 cont. Turbot in Subarea 4. Internal consistency of the three tuning indices available for the assessment: BTS-ISIS, SNS (previous page) and NL beam trawl LPUE (not used in the final assessment).


Figure 18.2.7. Landings (left) and stock (right) weight at age from observations (points) and modelled values (lines).


Figure 18.4.1 Turbot in Subarea 4. Key metrics from the assessment model results: spawner stock biomass (top left), mean fishing mortality (ages 2-6; top right), fishing mortality at age (bottom right) and recruitment (bottom left). The best estimate (solid line) and $95 \%$ confidence limits (dashed lines, shaded area) are plotted (except for F at age).


Figure 18.4.2 Turbot in Subarea 4. Retrospective analysis of the key metrics from the assessment model results: spawner stock biomass (top left), fishing mortality (top right) and recruitment (bottom). The current assessment (black line and shaded area for $95 \%$ uncertainty bounds) and five retrospective 'peels' (coloured lines) are shown.


Figure 18.5.1. Turbot in Subarea 4. Log catchability residuals for the landings. Red solid bubbles indicate that the model estimates higher values than observed, hollow blue bubbles indicate that the model estimates lower values than observed.


Figure 18.5.2. Turbot in Subarea 4. Log catchability residuals for the three tuning fleets: BTS-ISIS (top left), NL beam trawl LPUE (top right), and SNS (top). Red solid bubbles indicate that the model estimates higher values than observed, hollow blue bubbles indicate that the model estimates lower values than observed.


Figure 18.5.3. Turbot in Subarea 4. Leave-one-out (LOO) assessment results: spawner stock biomass (top left), fishing mortality (top right) and recruitment (bottom). The full assessment (black line and shaded area for $95 \%$ uncertainty bounds) and three alternative assessments, each leaving out one of the tuning indices: no SNS (purple), no BTS-ISIS (green) and no NL BT2 LPUE (pink).

## 19 Turbot (Scophthalmus maximus) in Division 3.a (Skagerrak and Kattegat)

In 2016, no advice was scheduled for this stock, which is under a biennial advice. So this section only reviews the latest information available ( 2015 catches, 2015 Quarter 3 and 2016 Quarter 1 surveys).

The general perception is that landings have increased, and that the latest survey estimates are also somehow higher than in the past few years. But these estimates are very noisy with large variations from year to year. WGNSSK considered that the perception of the stock had not changed and did not re-open the advice.

Discards ratio in Kattegat have increased in 2015.

### 19.1 Management regulations

There are no TACs in place for turbot in area 3.a. So far, no analytical assessments leading to fisheries advice have been carried out for turbot in 3.a by ICES, but some work is ongoing, testing various methods for the assessment of data limited stocks. No precautionary reference points have been proposed, and no management plans are in place for this stock.

There is no official EC minimum landing size.

### 19.2 Fisheries data

In 3.a, a target fisheries for turbot probably only occurred when the stock was large (i.e. before 1960s; Cardinale et al., 2009), while today turbot is only caught as by-catch in the trawl and gillnet fisheries. Table 19.1 and Figure 19.1 summarize turbot landings in ICES area 3.a. Over the period 1950 - 2015, total landings (3.a) ranged from 64 t to 736 t per year, with the lowest landings during the end of 1960's and the beginning of the 1970s, and the highest peaks in 1977 and in the early nineties. In the last decade, the total landings of turbot in 3.a had declined from around 350 t pr year to around 100 t per year, but landings in 2015 were higher ( 175 tonnes).

2015 catch data for turbot-kask were uploaded into InterCatch, according to the specification of the data call. This allowed compiling information by area and metier. Length-based information was provided (mainly for Kattegat), but no ages.

Discard ratios were provided for strata summing up to $72 \%$ of the reported landings ( $71 \%$ in Skagerrak, $79 \%$ in Kattegat). For those strata where information exist, discards ratios were estimated at $31 \%$ of catches in the Kattegat, but only $4 \%$ in the Skagerrak. Overall, this discard ratio is almost twice as high as last year in Kattegat ( $16 \%$ in 2014), but at the same level as before for the Skagerrak.
The raising of discards was performed by groups of métiers: all passive gears together (discards ratio close to zero), all trawled gears with mesh size $>=120 \mathrm{~mm}$ together (medium discards ratio). After raising, the discard ratio for the entire stock area was estimated at $9.1 \%$, which is an slightly increasing trend ( $6 \%$ in $2013,8 \%$ in 2014) (Table 19.2).

### 19.3 Survey data, recruit series and analysis of stock trends

Two survey series catching turbot are available: the International Bottom Trawl Survey (IBTS), with two research vessels (Argos and Dana), and the Baltic International Trawl

Survey (BITS) with the Danish vessel Havfisken (KASU survey). But since the initial investigations of ICES WGNEW, only the Havfisken trawl survey (BITS) is used to derive an index of abundance of turbot in 3.a. The estimated CPUE is computed in $n /$ hour, and a biomass proxy is calculated using a fixed length-weight relationship from www.fishbase.org ( $a=0.00802, b=3.260$ ).

Indices are noisy (Table 19.3), but in both surveys the last CPUE estimate (2016 for BITS Q1, and 2015 for BITS Q4) is higher than the last two years. In all four cases (for the two seasons and both in biomass and in number), the $2: 3$ ratio is above 1 . The index used for advice is BITS Q1 in number, and for this one the DLS 2:3 ratio is at 1.196, indicating an increase of $20 \%$ in the average of the last two years compared to the previous 3 years.

The length frequency distribution estimated from BITS surveys and aggregated every 5 years does not show major variations since 1990. Most of the fish caught are under 30 cm (Figure 19.2).

### 19.4 Re-opening of advice

According to the $2: 3$ rule, the advice in 2016 is more optimistic than the advice in 2015, which called for a reduction of $20 \%$. Nevertheless, WGNSSK notes that the indices available are quite noisy. The average number of fish caught per hour is very small, and with relatively few hauls in the survey. Therefore small variations in the survey catches lead to large differences when expressed in relative changes from year to year. In reality, the BITS Q1 index in number fluctuates around the mean, with a slightly decreasing trend over the time. Therefore, the relatively higher value in 2016 cannot be considered as an increase of the stock at this stage. A few more years would be necessary to confirm this point for being either a noisy estimate or indicative of a real trend.

On this basis, WGNSSK decided not to re-open the advice.

### 19.5 Biological sampling

WGNEW (2013) noted that turbot is classified as a Group 2 species under the DCF, and detailed some issues limiting the ability of member states to collect biological information for that stock.

### 19.6 Data recommendations

WGNEW (2013) formulated a number of recommendations for improving the biological knowledge on this stock:

In order to meet the DCF-requirements for sampling of biological parameters for turbot in the Kattegat- Skagerrak, the following countries could be valid candidates to fill current data gaps, according to their importance in turbot fisheries;

- Denmark in the Kattegat-Skagerrak
- Sweden in Kattegat-Skagerrak


## General recommendations

- EU to upgrade turbot from Group 2 to Group 1, forcing relevant Member States to collect biological information on a yearly basis
- Relevant Member States to include market sampling for turbot in their National Proposals, thus generating the required funds through the DCF.


### 19.7 References

ICES. 2013. Report of the Working Group on Assessment of New MoU Species (WGNEW), 18 22 March 2013, ICES HQ, Copenhagen, Denmark. ACOM .

Cardinale, M., Linder, M., Bartolino, V., Maiorano, L., and Casini, M. 2009. Conservation value of historical data: Reconstructing stock dynamics of turbot during the last century in the Kattegat-Skagerrak. Marine Ecology Progress Series, 386: 197-206.

Table 19.1. Turbot in 3.a: Official landings by country from 1950 to 2014.

| Year | BEL | DEU | DNK | GBR | NLD | NOR | SWE | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0 | 13 | 212 | 0 | 0 | 1 | 73 | 299 |
| 1951 | 0 | 6 | 191 | 0 | 0 | 6 | 62 | 265 |
| 1952 | 0 | 6 | 114 | 0 | 0 | 3 | 58 | 181 |
| 1953 | 0 | 4 | 80 | 0 | 0 | 4 | 51 | 139 |
| 1954 | 0 | 0 | 78 | 0 | 0 | 1 | 61 | 140 |
| 1955 | 0 | 4 | 77 | 0 | 0 | 0 | 49 | 130 |
| 1956 | 0 | 7 | 75 | 0 | 0 | 0 | 41 | 123 |
| 1957 | 0 | 3 | 108 | 0 | 0 | 0 | 30 | 141 |
| 1958 | 0 | 7 | 112 | 0 | 0 | 0 | 41 | 160 |
| 1959 | 0 | 6 | 132 | 0 | 0 | 3 | 43 | 184 |
| 1960 | 0 | 11 | 115 | 0 | 0 | 2 | 46 | 174 |
| 1961 | 0 | 4 | 130 | 0 | 0 | 0 | 45 | 179 |
| 1962 | 0 | 5 | 157 | 0 | 0 | 0 | 0 | 162 |
| 1963 | 0 | 4 | 124 | 0 | 0 | 0 | 0 | 128 |
| 1964 | 0 | 5 | 89 | 0 | 0 | 0 | 0 | 94 |
| 1965 | 0 | 6 | 79 | 1 | 0 | 0 | 0 | 86 |
| 1966 | 0 | 2 | 104 | 0 | 0 | 0 | 0 | 106 |
| 1967 | 0 | 4 | 68 | 1 | 0 | 0 | 0 | 73 |
| 1968 | 0 | 0 | 64 | 0 | 0 | 0 | 0 | 64 |
| 1969 | 0 | 1 | 75 | 0 | 0 | 0 | 0 | 76 |
| 1970 | 0 | 1 | 76 | 0 | 0 | 0 | 0 | 77 |
| 1971 | 0 | 1 | 100 | 0 | 0 | 0 | 0 | 101 |
| 1972 | 0 | 2 | 130 | 0 | 0 | 0 | 0 | 132 |
| 1973 | 0 | 2 | 98 | 0 | 0 | 0 | 0 | 100 |
| 1974 | 0 | 1 | 116 | 0 | 0 | 0 | 0 | 117 |
| 1975 | 0 | 2 | 167 | 0 | 7 | 0 | 7 | 183 |
| 1976 | 7 | 2 | 178 | 0 | 190 | 0 | 6 | 383 |
| 1977 | 7 | 4 | 331 | 0 | 389 | 0 | 5 | 736 |
| 1978 | 2 | 4 | 327 | 0 | 186 | 0 | 6 | 525 |
| 1979 | 8 | 0 | 307 | 0 | 87 | 0 | 4 | 406 |
| 1980 | 7 | 0 | 205 | 1 | 14 | 0 | 6 | 233 |
| 1981 | 2 | 0 | 183 | 2 | 12 | 0 | 8 | 207 |
| 1982 | 1 | 0 | 164 | 1 | 9 | 0 | 7 | 182 |
| 1983 | 4 | 0 | 171 | 0 | 24 | 0 | 10 | 209 |
| 1984 | 0 | 0 | 176 | 0 | 0 | 0 | 12 | 188 |
| 1985 | 1 | 0 | 224 | 0 | 0 | 0 | 16 | 241 |
| 1986 | 2 | 0 | 180 | 0 | 0 | 0 | 11 | 193 |
| 1987 | 5 | 0 | 147 | 0 | 0 | 0 | 9 | 161 |
| 1988 | 2 | 0 | 115 | 0 | 11 | 0 | 10 | 138 |
| 1989 | 2 | 0 | 173 | 0 | 0 | 0 | 9 | 184 |
| 1990 | 5 | 0 | 363 | 0 | 0 | 0 | 18 | 386 |
| 1991 | 4 | 0 | 244 | 0 | 0 | 7 | 21 | 276 |
| Year | BEL | DEU | DNK | GBR | NLD | NOR | SWE | Total |


| Year | BEL | DEU | DNK | GBR | NLD | NOR | SWE | TOTAL |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1992 | 4 | 0 | 278 | 0 | 0 | 8 | 19 | 309 |
| 1993 | 3 | 0 | 336 | 0 | 0 | 10 | 0 | 349 |
| 1994 | 2 | 0 | 313 | 0 | 0 | 15 | 22 | 352 |
| 1995 | 4 | 0 | 268 | 0 | 0 | 17 | 11 | 300 |
| 1996 | 0 | 0 | 185 | 0 | 0 | 13 | 11 | 209 |
| 1997 | 0 | 0 | 200 | 0 | 0 | 9 | 11 | 220 |
| 1998 | 0 | 0 | 148 | 0 | 0 | 7 | 8 | 163 |
| 1999 | 0 | 0 | 139 | 0 | 0 | 10 | 6 | 155 |
| 2000 | 0 | 0 | 180 | 0 | 0 | 6 | 6 | 192 |
| 2001 | 0 | 0 | 227 | 0 | 0 | 8 | 3 | 238 |
| 2002 | 0 | 0 | 205 | 0 | 0 | 11 | 5 | 221 |
| 2003 | 0 | 0 | 128 | 0 | 13 | 14 | 4 | 159 |
| 2004 | 0 | 0 | 119 | 0 | 14 | 7 | 7 | 147 |
| 2005 | 0 | 0 | 108 | 0 | 7 | 6 | 6 | 127 |
| 2006 | 0 | 1 | 95 | 0 | 8 | 8 | 9 | 121 |
| 2007 | 0 | 1 | 138 | 0 | 15 | 7 | 12 | 173 |
| 2008 | 0 | 1 | 121 | 0 | 4 | 6 | 11 | 143 |
| 2009 | 0 | 1 | 94 | 0 | 2 | 6 | 17 | 120 |
| 2010 | 0 | 0 | 72 | 0 | 6 | 4 | 13 | 95 |
| 2011 | 0 | 1 | 78 | 0 | 0 | 7 | 13 | 99 |
| 2012 | 0 | 0 | 168 | 0 | 0 | 8 | 14 | 189 |
| 2013 | 0 | 0 | 91 |  |  | 5 | 15 | 111 |
| 2014 | 0 | 1 | 94 | 0 | 2 | 6 | 17 | 120 |
| 2015 | 0 | 0 | 135 | 0 | 20 | 8 | 11 | 175 |
|  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |

Table 19.2. Turbot in 3.a: Landings and discards (in kg ) after raising in InterCatch.

|  | Discards | Landings | Grand Total | Discard <br> Ratio |
| :--- | :--- | :--- | :--- | :--- |
| 2013 | 7365.87 | 112960 | 120325.87 | $6.1 \%$ |
| 3.aN | 1904.79 | 78830 | 80734.79 | $2.4 \%$ |
| 3.aS | 5461.08 | 34130 | 39591.08 | $13.8 \%$ |
| 2014 | 10508.24 | 120240.6 | 130748.84 | $8.0 \%$ |
| 3.aN | 2712.3 | 80968.9 | 83681.2 | $3.2 \%$ |
| $3 . \mathrm{aS}$ | 7795.94 | 39271.7 | 47067.64 | $16.6 \%$ |
| 2015 | 18274 | 183501.8 | 201775.8 | $9.1 \%$ |
| 3.aN | 4639 | 145083.9 | 149722.9 | $3.1 \%$ |
| 3.aS | 13635 | 38417 | 52052 | $26.2 \%$ |

Table 19.3. Turbot in 3.a: Average CPUE ( $\mathbf{n} / \mathrm{h}$ ) estimated from BITS (KASU) surveys for quarter 1 and quarter 4 between 1996 and 2016, and DLS calculations using 2:3 rule

|  | Q1 | Q1 | Q4 | Q4 |
| :---: | :---: | :---: | :---: | :---: |
| Year | wt | number | wt | number |
| 1992 |  |  |  |  |
| 1993 |  |  | 1.294 | 2.000 |
| 1994 |  |  | 0.744 | 2.000 |
| 1996 | 0.860 | 1.955 | 0 | 0.000 |
| 1997 | 0.557 | 0.422 |  |  |
| 1998 | 1.033 | 2.500 |  |  |
| 1999 | 1.121 | 1.227 | 0.962 | 1.727 |
| 2000 | 0.430 | 1 | 0.207 | 0.667 |
| 2001 | 0.689 | 1.818 | 0.597 | 1.185 |
| 2002 | 0.254 | 0.621 | 0.683 | 1.515 |
| 2003 | 0.732 | 1.848 | 0.236 | 0.515 |
| 2004 | 0.445 | 1.030 | 4.088 | 2.258 |
| 2005 | 0.393 | 0.894 | 0.636 | 1.197 |
| 2006 | 0.884 | 1.227 | 0.362 | 0.470 |
| 2007 | 0.361 | 0.894 | 0.427 | 1.000 |
| 2008 | 0.809 | 1.788 | 0.517 | 0.867 |
| 2009 | 0.336 | 0.633 | 0.303 | 0.606 |
| 2010 | 0.717 | 1.533 | 0.783 | 2.233 |
| 2011 | 0.400 | 0.803 | 0.437 | 0.712 |
| 2012 | 0.416 | 0.591 | 0.304 | 0.742 |
| 2013 | 0.944 | 1.455 | 0.216 | 0.545 |
| 2014 | 0.190 | 0.482 | 0.227 | 0.377 |
| 2015 | 0.458 | 0.652 | 0.926 | 1.167 |
| 2016 | 1.200 | 1.364 |  |  |
|  |  |  |  |  |
| 2 yrs average | 0.829 | 1.008 | 0.577 | 0.772 |
| 3 previous | 0.517 | 0.843 | 0.319 | 0.666 |
| 2:3 | 1.605 | 1.196 | 1.807 | 1.159 |



Figure 19.1. Turbot in 3.a: official landings by country from 1950 to 2015.


Figure 19.2. Turbot in 3.a. Trend in cpue (in $\mathrm{kg} / \mathrm{h}$, left or in $\mathrm{n} / \mathrm{h}$, right) estimated from BITS surveys between 2004 and 2016 in quarter 1 (top) and quarter 4 (bottom).


Figure 19.7. Turbot in 3.a. Length frequency distribution derived from BITS surveys in quarter 1 and aggregated every 5 years.


Figure 19.8. Turbot in 3.a. Length frequency distribution derived from BITS surveys in quarter 4 and aggregated every 5 years.

## 20 Brill (Scophthalmus rhombus) in Subarea 4 and Divisions 3.a and 7.d-e (North Sea, Skagerrak and Kattegat, English Channel)

Brill (Scophthalmus rhombus) was assessed in the Working Group on Assessment of New MoU Species until 2013 (ICES, 2013a). Because only official landings and survey data were available, brill in the Greater North Sea was defined as a category 3 stock according to the ICES guidelines for data limited stocks (ICES, 2012a). WGNSSK refined the WGNEW-advice in this final year (ICES, 2013b), and biennial advice was given in 2013 (ICES, 2013c) on the basis of LPUE-trends of the Dutch beam trawl fleet (vessels > 221 kW ). In 2014-2016, the stock was included in the data call for the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK) and the biennial advice was evaluated in this group. 2015 and 2016 were the first years in which landings and discards by metier were requested from all countries contributing to the North Sea brill fishery through InterCatch.

### 20.1 General

### 20.1.1 Biology and ecosystem aspects

Brill is a shallow-water flatfish mainly found in areas close inshore. It prefers sandy bottoms, but can sometimes also be found on gravel and muddy grounds. Its vertical distribution ranges from 4 meters to 73 meters, although small juvenile fish are often common in sand shore pools. Mature brill are rarely observed inshore, whereas immature specimens are often caught near the coast and even in estuaries.

The distribution of brill in the North Eastern Atlantic ranges along the European coastline from $64^{\circ} \mathrm{N}$ (the Lofotes) down to $30^{\circ} \mathrm{N}$, extending into the Mediterranean and even into the Black Sea (Nielsen, 1986). Brill is also found in the Skagerrak, the Kattegat, and small quantities in the Baltic Sea. The western limit of its distribution area is reached in southern Iceland.

The feeding habits of this species closely resemble those of turbot and were extensively reviewed by de Groot (1971) and Wetsteijn (1981). The pelagic larvae feed primarily on copepod nauplii, decapod and mollusc larvae. With increasing size, this diet gradually changes from larger invertebrate prey and larvae of several fish species to small fish. Larger brill ( $>40 \mathrm{~cm}$ ) are primarily piscivorous.
More information on the biology of brill can be found in Annex 5 of WGNEW 2010 (ICES, 2010).

### 20.1.2 Stock identity and possible assessment areas

The oldest study that could be found containing information on the genetic structure of brill was carried out by Blanquer et al. (1992), using allozyme electrophoresis. No genetic differentiation could be found between Atlantic and Mediterranean populations, suggesting that there are also very low levels of differentiation in brill from different areas.
In the EU funded study on 'Stock discrimination in relation to the assessment of the brill fishery' the following was concluded (Delbare and De Clerck, 1999): "As a final conclusion, biological parameters (composition of Belgian brill landings, growth rate and reproduction characteristics) and the sequencing of the D-loop resulted in insignificant differences between brill from the different areas. Therefore, arguments favour
the hypothesis that brill from the NE Atlantic might be considered to be only one population: the North-eastern Atlantic brill population. Further research on spawning areas and migration through respectively egg surveys and tagging experiments, could generate valuable information about (sub-)population structures of brill throughout its entire distribution area. Therefore it is advisable to extend the sampling area to the Mediterranean Sea and the Black Sea."

Recently, the genetic structure of brill over its entire distribution area has being characterized by Vandamme (2014). Genetic variation was found to be of mean to high levels, but the results show almost no differentiation between potential biological populations and/or management units. Therefore, we still feel confident in treating brill in 3.a, 4 and 7.de as a single stock, which could potentially have an even wider geographical spread.

Further research on brill spawning areas (egg surveys), and of migration of adult (tagging experiments) and especially immature brill (tagging experiments and genetic analysis of the immature population components) could still generate valuable information about (sub-)population structure of brill throughout its entire distribution area.

More information on the delineation of potential brill stocks can be found in Annex 5 of WGNEW 2010 (ICES, 2010).

### 20.1.3 Management regulations

Although several EC regulations affect the flatfish fisheries in the North Sea (e.g. effort restrictions, minimum mesh sizes), no explicit management objectives have been defined for the stock of brill in the Greater North Sea, and no management plans are in place. However, for the EU-waters in Division 2.a and Subarea 4, precautionary TACs have been defined for brill and turbot (combined). It is unclear how the quantitative single species advices for turbot and brill are used to formulate a combined TAC, that belongs entirely to the EU-fisheries. A historical overview is presented in the table below.

Historical overview of combined TACs for brill Scophthalmus rhombus and turbot Scophthalmus maximus in Division 2.a and Subarea 4

| YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| TAC | 9000 | 9000 | 6750 | 5738 | 4877 | 4550 | 4323 | 4323 | 5263 |
| Year | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |  |
| TAC | 5263 | 5263 | 4642 | 4642 | 4642 | 4642 | 4642 | 4488 |  |

No restriction on the minimum length for landing brill is imposed by the EC. Some authorities have however installed Minimum Landing Sizes (MLS) for brill. The most frequently applied MLS is 30 cm (e.g., in Belgium, the Baltic, the English Sea Fisheries District Cornwall, ...).

### 20.2 Fisheries data

### 20.2.1 Landings

Tables 20.1-3 summarize the official brill landings by country for Subarea 4, Division 3.a and Divisions 7.d-e respectively (Source: ICES Fishstat). The total international landings can be consulted in Table 20.4 and Figure 20.1. Over the period 1950-1970, total landings ranged from 582 t to 947 t per year, followed by a gradual increase to 2121 t in 1977. During 1978-2014, total landings varied between 1517 t (in 1980) and

3141 t (in 1993). In 2000-2014, annual total landings fluctuated around an average of 2112 t (range: $1781 \mathrm{t}-2409 \mathrm{t}$ ). The higher value of The North Sea (4) accounts for the major part of these landings (Figure 20.2), on average generating $68 \%$ of the totals over the time series (range: 50-86\%). In 2015, landings increased to the third highest value in the time series ( 2489 t ). The English Channel and the Skagerrak are responsible for average contributions to the international brill landings of $19 \%$ and $13 \%$ respectively. The Skagerrak was responsible for a higher relative importance in the total landings during the first two decades of the time series, and the English Channel has gained importance since the late seventies. No trend towards a higher or lower mean relative contribution of a certain Subarea or Division is apparent in the data for the more recent years. It is however possible that these trends (or lack thereof) are influenced by incomplete statistics for the early part of the time series.

Uptake percentages for brill in the Greater North Sea assessment area cannot be reliably calculated, as the TAC is set combined with turbot. Additionally, there is a mismatch between the assessment and the management areas, as the TAC is set for Subareas 4 and Division 2.a.

More details on the Belgian, Dutch, French and UK fisheries catching brill, and information on length- and age-distributions of Belgian brill landings can be found in Annex 5 of WGNEW 2010 (ICES, 2010).

### 20.2.2 Discards

Due to its high value and the absence of a European Minimum Landing Size, brill is not expected to be discarded easily by fishermen catching the species as long as the quota have not been fully taken. The fact that the species is characterized by a fast growth, quickly reaching commercially interesting lengths (unfortunately at relatively young ages and while still immature), explains why smaller individuals are rather rare in commercial catches, contributing to the low numbers of discards. Therefore, earlier evaluations resulted in the labelling of this stock as one with negligible discards, and landings were considered to be a reliable proxy for total catch. The amount of discarding of brill was not thought to be a substantial problem for the assessments of the state of the species's stocks in terms of data completeness. From a biological perspective however, it's a very different story, as most of the discarded fish have not reached sexual maturity yet, and as such have not had the chance to reproduce and contribute to the future generations and future fisheries.

In 2014, discard rates and/or discard data that were raised to fleet levels were available for the first time through InterCatch, for some countries participating in the brill fishery. However, these were not analyzed or incorporated in the assessment, or to top up the landings in order to issue a catch advice, as 2014 was an update year. Under the 2015 data call, 8 countries (Belgium, Denmark, France, Germany, Ireland, The Netherlands, Sweden and the United Kingdom) were expected to upload quarterly discard data by métier and Division in InterCatch, for the years 2012-2014. The response to this data call can be evaluated as very good for this stock. Five countries delivered data for all three years (Belgium, Germany, Netherlands, Norway - that was not in the call, and United Kingdom - both England and Scotland), three countries for 2013 and 2014 (Denmark, France and Sweden), and Ireland for 2014. All subdivisions and the main métiers catching brill (gear types: TBB, OTB, GNS, GTR) were covered in this way, and quar-ter/country-combinations for which discard information was lacking could be easily filled by allocating discard rates of similar/identical métiers or quarters. The 2016 data call listed the same data demands, and also the response of the different countries can be evaluated as good (both quantitatively and qualitatively).

The resulting overall overview of the raised discard rates in 2012-2015 is shown in Table 20.5, and broken down by country and Subarea/Division for the years 2014-2015 (years with the most complete coverage) respectively in Tables 20.6-7. The overall discard rates in these years ( $7 \%$ in 2012, $4 \%$ in $2013,8 \%$ in $2014,9 \%$ in 2015) show no trend (Table 20.5), but may have gone up compared to the years prior to 2012. The overall average discard rate over 2012-2014 is 7\%. A comparison can be made with the discard rate time series 2008-2012 for the Belgian TBB_DEF in WGNEW 2013 (ICES, 2013.a). At that time, $7 \%$ (in 2011 in 4) was the highest value ever documented in this programme.

The overview by country (Table 20.6, listing discard rates for 2014 and 2015) shows discard rates that are well above the average for Sweden (29-35\%) and Denmark (21$22 \%$ ) in both years, corresponding to the higher discard rates in the North of the assessment area (up to $38 \%$ in 3.a; Table 20.7). These higher numbers in the North are largely caused by gillnet and trammel net fisheries taking place there. Remarkable is that the high discard rate of $16 \%$ for Germany in 2014 dropped to only $1 \%$ in 2015 , and that the rather low discard rate of $4 \%$ for France in 2014 increased to $24 \%$ in 2015.

Length and/or age distributions of brill discards were not requested in the 2015 and 2016 data calls. Details on the numbers at length discarded per hour in the Dutch beam trawl fleet (North Sea) can be found in Annex 5 of WGNEW 2010 (ICES, 2010), and length distributions for the Belgian beam trawl fleet (4, 7.d and 7.e, for the years 2008 - 2012) are presented in WGNEW 2013 (ICES, 2013.a).

### 20.3 Tuning series

### 20.3.1 Survey Data

## General

Catches of brill are generally very low on surveys. These low catch numbers very often result in an underrepresentation of some year- or length-classes (mainly the older or bigger ones), leading to a poor quality of the resulting survey abundance series and indices, and poor agreement among different surveys.

WGNEW 2012 (ICES, 2012b) tested four surveys for their potential use in describing stock trends of brill in the greater North Sea. Three of these surveys take place in the North Sea (IBTS_TRI_Q1, BTS_TRI_Q3 and BTS_ISI_Q3) and one in the English Channel (CGFS_Q4). Time series of total numbers of brill caught by the three North Sea surveys and the Channel are depicted in WGNEW 2012 (ICES, 2012b), but only the BTS_ISI_Q3 was found to catch a sufficient number of individuals to be useful in the context of evaluating stock trends of North Sea brill. WGNEW 2013 and the following WGNSSK-meetings did not go into these surveys again, with exception for the BTS_ISI_Q3 and BITS_HAF_Q1\&4 that were updated because of their use as indicators in the advice that was issued in 2013 (ICES, 2013c) in the North Sea and the Skagerrak respectively. Plots and tables for these surveys were also updated during WGNSSK 2016.

## North Sea (Subarea 4)

The abundance indices (numbers per hour) for brill in the BTS_ISI_Q3 in 4 are spatially plotted per rectangle in Figure 20.3 and over time in Figure 20.4 and Table 20.8. These seem to illustrate a recovery of the species in 4 since 2009 after a period of consistent lower catches during 2001-2008, followed by a drop in abundance in 2012-2013, a steep
increase in 2014 and again a drop in 2015. The inter-annual variation between all other years is so big that no real trend is apparent over the entire time series. Therefore, the lower catches per hour in $2012(1.04 / \mathrm{hr})$ in comparison with the higher values in the three preceding years ( $1.42-2.41 / \mathrm{hr}$ ), were not yet considered to represent an alarming signal by WGNEW 2013 (ICES 2013.a). The confirmation of the low abundance by the survey in 2013 ( $0.76 / \mathrm{hr}$, even lower than in 2012 and the $6^{\text {th }}$ lowest value in the time series) raised some concerns during WGNSSK 2013 (ICES 2013b), that appears to be no longer relevant as the abundance index jumped up to $3.04 / \mathrm{hr}$ in 2014 ( $3^{\text {rd }}$ highest value of the time series). The lower value in $2015(1.84 / \mathrm{hr})$ is still higher than the long term average ( $1.55 / \mathrm{hr}$ ).

The corresponding ALK, length distributions (per 5 years) and length-at-maturity are illustrated in Figures 20.5-7. These show that mainly brill of ages 1-2 and lengths of $20-45 \mathrm{~cm}$ are caught in this survey and that no obvious shifts in length distributions are apparent over the time series (1987-2015). All brill under 30 cm were immature, and all above 40 cm were mature, with a mix of mature and immature individuals between 30 and 40 cm .

## Skagerrak/Kattegat (Division 3.a)

Data on brill from the Danish BITS-survey in the Kattegat (BITS_HAF_Q1\&4) were analysed separately for the two quarters in which this survey runs by WGNEW 2013 (ICES, 2013.a), revealing almost identical patterns for Q1 and Q4. Therefore, it was decided to combine the data from both quarters for the evaluation of the brill substock in 3.a, and only the results of this combined analysis are updated and presented in this report. The fact that this survey only covers the Kattegat (3.aS) and not the Skagerrak (3.aN) was not considered to be a problem by WGNEW 2013 as the deeper northern waters don't harbour important numbers or densities of brill, that generally prefers more shallow waters.

The abundance indices (numbers per hour) for brill in the BITS_HAF_Q1\&4 are spatially plotted per rectangle in Figure 20.8 and over time in Figure 20.9 and Table 20.9. These illustrate a period with higher catches (2006-2011) after a period of consistent lower catches (1996-2005). In 2012, the numbers caught per hour dropped to the level of 2004-2005 again but given the noise in the data (large inter-annual variations) it was considered to be preliminary to interpret this as a sign of a decreasing stock. However, as in the survey used as an indicator for brill in the North Sea, the lower abundance of $2012(2.27 / \mathrm{hr})$ in 3.a was also followed by an even lower abundance in $2013(2.13 / \mathrm{hr})$, and a steep increase in 2014 (3.86/hr) up to the highest value ever documented in this survey in 2015 (4.47/hr). Although the survey index values are generally higher in 3.a compared to 4, the trends are remarkably similar in the past few years, except for 2015 (decrease in 4, increase in 3.a).

The corresponding length distributions (per 5 years) for the BITS_HAF_Q1\&4 in 3.a are shown in Figure 20.10. As in Subarea 4, no alarming shifts in length distributions (no obvious loss of larger/older individuals from the population) are apparent over the time series (1996-2014). A much bigger overlap in length between the immature and mature stages compared to the North Sea, with mature individuals of lengths lower than 20 cm , was documented in WGNEW 2013 (ICES 2013.a). This illustrates the general phenomenon of slower growth at higher lattitudes that was also published for brill by Delbare \& Declerck (1999), that didn't include the Skagerrak/Kattegat in their overview.

## English Channel (Divisions 7.d-e)

Unfortunately, no useful survey index could be identified for the evaluation of the brill substock in the English Channel during previous WGNEW meetings (ICES, 2010, 2012b, 2013.a).

### 20.3.2 Commercial LPUE series

Although the survey indices presented above are useful indicators when evaluating the state of the brill stock in (parts of) the Greater North Sea area, the spatial coverage of both surveys was evaluated as insufficiently spanning the stock area, and the catches too low, to use these surveys as a basis for catch advice, by previous WGNEW and WGNSSK meetings.
A corrected Landings Per Unit of Effort (LPUE) series from the Dutch beam trawl fleet > 221 kW was presented to and discussed for the first time during WGNEW 2013 (see ICES, 2013.a for interpretation), and has been used as the basis for the advice since. These LPUE's were standardized for engine power and corrected for targeting behavior in a way similar to the one used to analyze commercial LPUE data for North Sea plaice. The standardization for engine power is relevant as trawlers are likely to have higher catches with higher engine powers, as they can trawl heavier gear or fish at higher speeds. The correction for targeting behavior relies on reducing the effects of spatial shifts in fishing effort by calculating the fishing effort by ICES rectangle and subsequently averaging these over the entire fishing area. More information on the data that were used (EU logbook auction data and market sampling data), the calculation of the LPUEs, the standardization of engine power, the correction for targeting behavior and the results can be found in van der Hammen et al. (2011).

In 2015, a revised LPUE series was delivered to the Working Group, containing fewer years than in the past. Both the old (delivered to WGNSSK 2014) and new (delivered to WGNSSK 2015) series were presented in the corresponding report (ICES 2015). Given the facts that the majority of the brill landings from the greater North Sea originate from Subarea 4, and that around $70 \%$ (on average) of these are landed by the Netherlands, this LPUE series may be considered a more reliable time series when evaluating the stock trend of brill in the Greater North Sea. Differences between the two series are discussed in the report of WGNSSK 2015 (ICES 2015), in 2016 only the new series - the one used for the advice - is presented. This series showed a consistently increasing LPUE (kg/day) up to 2012, dropping slightly over 2013-2014 (6\% decrease between 2010-2012 and 2013-2014) but increasing again in 2015. As a result, applying the 2:3 rule leads to an increase of 1\% between 2011-2013 and 2014-2015.

### 20.4 Biological sampling and population biology parameters

No new information was obtained compared to previous WGNEW and WGNSSK reports.

### 20.5 Analyses of stock trends and potential status indicators

So far, no analytical assessments leading to fisheries advice have been carried out for brill in the Greater North Sea by ICES. In the absence of collated and analyzed biological data, Category 3 of the ICES Data Limited Stocks Methodology (ICES, 2012a) is currently the highest attainable category for this stock. Method 3.2.0 specifies that catch advice can be derived from the survey-adjusted status-quo catch in situations where there are survey data on abundance (e.g. CPUE over time), but survey-based proxies
for MSY $\mathrm{B}_{\text {trigger }}$ and F values are not known. Also other indicators of stock size can be used.

WGNEW and WGNSSK tested several surveys for their information content regarding brill in the Greater North Sea over the last few years, and decided to retain only the BTS_ISI_Q3 in 4 and the BITS_HAF_Q1\&4 in 3.a as useful indicators for this stock, and to use the commercial LPUE series from the Dutch beam trawl fleet $>221 \mathrm{~kW}$ as a more reliable time series and basis for the advice.

As brill in the Greater North Sea is a stock for which biennial catch advice was issued in 2015 (valid for 2016 and 2017), and the perception of the stock hasn't changed since, the WG decided not te reopen this advice and stick to the advice issued in 2015.

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Table 20.1 Landings ( $\mathbf{t}$ ) of brill Scophthalmus rhombus in Subarea IV (North Sea) by country, over the period 1950-2015 (Source: ICES Fishstat).

| Year | BEL | GER | DNK | FRA | GBR | NLD | NOR | SWE | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 34 | 0 | 39 | 0 | 183 | 108 | 1 | 19 | 384 |
| 1951 | 23 | 0 | 53 | 0 | 322 | 93 | 1 | 19 | 511 |
| 1952 | 21 | 0 | 65 | 0 | 350 | 117 | 3 | 9 | 565 |
| 1953 | 23 | 0 | 49 | 0 | 376 | 130 | 0 | 11 | 589 |
| 1954 | 19 | 0 | 53 | 0 | 330 | 106 | 14 | 7 | 529 |
| 1955 | 23 | 0 | 51 | 0 | 357 | 137 | 3 | 0 | 571 |
| 1956 | 28 | 0 | 47 | 0 | 276 | 156 | 0 | 9 | 516 |
| 1957 | 32 | 0 | 27 | 0 | 247 | 154 | 0 | 8 | 468 |
| 1958 | 43 | 0 | 42 | 0 | 223 | 162 | 0 | 10 | 480 |
| 1959 | 41 | 0 | 30 | 0 | 219 | 125 | 0 | 9 | 424 |
| 1960 | 55 | 0 | 37 | 0 | 235 | 150 | 1 | 8 | 486 |
| 1961 | 102 | 0 | 40 | 0 | 264 | 166 | 0 | 9 | 581 |
| 1962 | 97 | 0 | 42 | 0 | 238 | 214 | 0 | 0 | 591 |
| 1963 | 79 | 0 | 59 | 0 | 307 | 175 | 0 | 0 | 620 |
| 1964 | 79 | 0 | 46 | 0 | 161 | 279 | 0 | 0 | 565 |
| 1965 | 71 | 0 | 56 | 0 | 127 | 281 | 0 | 0 | 535 |
| 1966 | 100 | 0 | 63 | 0 | 119 | 264 | 0 | 0 | 546 |
| 1967 | 138 | 0 | 29 | 0 | 105 | 137 | 0 | 0 | 409 |
| 1968 | 152 | 0 | 43 | 0 | 110 | 274 | 0 | 0 | 579 |
| 1969 | 145 | 0 | 47 | 0 | 102 | 364 | 0 | 0 | 658 |
| 1970 | 114 | 0 | 42 | 0 | 76 | 386 | 0 | 0 | 618 |
| 1971 | 187 | 0 | 72 | 0 | 94 | 720 | 0 | 0 | 1073 |
| 1972 | 213 | 0 | 65 | 0 | 51 | 665 | 0 | 0 | 994 |
| 1973 | 185 | 0 | 55 | 0 | 39 | 710 | 0 | 0 | 989 |
| 1974 | 135 | 0 | 68 | 0 | 44 | 905 | 0 | 0 | 1152 |
| 1975 | 164 | 0 | 76 | 13 | 44 | 925 | 0 | 0 | 1222 |
| 1976 | 148 | 0 | 65 | 10 | 45 | 940 | 0 | 0 | 1208 |
| 1977 | 166 | 0 | 88 | 17 | 60 | 1079 | 0 | 0 | 1410 |
| 1978 | 175 | 0 | 123 | 26 | 84 | 967 | 0 | 0 | 1375 |
| 1979 | 188 | 0 | 154 | 10 | 103 | 908 | 0 | 0 | 1363 |
| 1980 | 129 | 0 | 104 | 8 | 45 | 747 | 0 | 0 | 1033 |
| 1981 | 148 | 0 | 66 | 5 | 42 | 957 | 0 | 0 | 1218 |
| 1982 | 182 | 0 | 53 | 11 | 41 | 1007 | 0 | 0 | 1294 |
| 1983 | 182 | 0 | 62 | 23 | 28 | 1153 | 0 | 0 | 1448 |
| 1984 | 190 | 0 | 73 | 30 | 29 | 1200 | 0 | 0 | 1522 |
| 1985 | 187 | 0 | 71 | 35 | 46 | 1370 | 0 | 0 | 1709 |
| 1986 | 131 | 0 | 76 | 4 | 46 | 950 | 0 | 0 | 1207 |
| 1987 | 140 | 0 | 50 | 17 | 48 | 715 | 0 | 0 | 970 |
| 1988 | 102 | 0 | 33 | 18 | 52 | 880 | 0 | 0 | 1085 |
| 1989 | 112 | 0 | 43 | 9 | 58 | 1080 | 0 | 0 | 1302 |
| 1990 | 168 | 0 | 139 | 24 | 82 | 480 | 0 | 0 | 893 |
| 1991 | 205 | 38 | 145 | 28 | 147 | 1111 | 8 | 0 | 1682 |


| YEAR | BEL | GER | DNK | FRA | GBR | NLD | NOR | SWE | TOTAL |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1992 | 203 | 59 | 77 | 34 | 218 | 1196 | 22 | 1 | 1810 |
| 1993 | 291 | 63 | 118 | 38 | 268 | 1647 | 14 | 0 | 2439 |
| 1994 | 208 | 90 | 109 | 28 | 235 | 1235 | 11 | 0 | 1916 |
| 1995 | 194 | 67 | 55 | 24 | 145 | 943 | 6 | 0 | 1434 |
| 1996 | 206 | 47 | 64 | 15 | 175 | 732 | 8 | 0 | 1247 |
| 1997 | 129 | 48 | 38 | 1 | 135 | 590 | 16 | 0 | 957 |
| 1998 | 160 | 58 | 58 | 11 | 172 | 808 | 16 | 0 | 1283 |
| 1999 | 161 | 51 | 91 | 0 | 156 | 805 | 16 | 0 | 1280 |
| 2000 | 167 | 77 | 93 | 16 | 141 | 998 | 16 | 0 | 1508 |
| 2001 | 182 | 66 | 67 | 12 | 158 | 1075 | 13 | 0 | 1573 |
| 2002 | 145 | 58 | 52 | 10 | 120 | 907 | 10 | 0 | 1302 |
| 2003 | 145 | 70 | 57 | 9 | 119 | 934 | 12 | 0 | 1346 |
| 2004 | 140 | 66 | 77 | 7 | 168 | 772 | 19 | 0 | 1249 |
| 2005 | 120 | 62 | 89 | 7 | 138 | 716 | 28 | 0 | 1160 |
| 2006 | 105 | 55 | 75 | 9 | 154 | 765 | 12 | 0 | 1175 |
| 2007 | 110 | 47 | 52 | 12 | 156 | 854 | 9 | 0 | 1240 |
| 2008 | 117 | 42 | 86 | 5 | 93 | 650 | 11 | 0 | 1004 |
| 2009 | 109 | 54 | 96 | 8 | 105 | 786 | 4 | 0 | 1162 |
| 2010 | 104 | 75 | 97 | 12 | 136 | 1072 | 4 | 0 | 1500 |
| 2011 | 101 | 57 | 122 | 13 | 137 | 1061 | 6 | 0 | 1497 |
| 2012 | 110 | 71 | 126 | 12 | 102 | 1084 | 7 | 0 | 1512 |
| 2013 | 100 | 63 | 123 | 10 | 117 | 972 | 4 | 0 | 1389 |
| 2014 | 98 | 69 | 96 | 9 | 116 | 811 | 9 | 4 | 1212 |
| 2015 | 149 | 115 | 122 | 7 | 136 | 1124 | 1 | 0 | 1655 |
|  |  |  |  |  |  |  |  |  |  |

Table 20.2 Landings ( $\mathbf{t}$ ) of brill Scophthalmus rhombus in Subdivision 3.a (Skagerrak) by country, over the period 1950-2015 (Source: ICES Fishstat).


| YEAR | BEL | GER | DNK | NLD | NOR | SWE | TOTAL |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1992 | 1 | 0 | 123 | 0 | 7 | 15 | 146 |
| 1993 | 2 | 0 | 184 | 0 | 10 | 16 | 212 |
| 1994 | 0 | 0 | 191 | 0 | 12 | 19 | 222 |
| 1995 | 0 | 0 | 124 | 0 | 13 | 14 | 151 |
| 1996 | 0 | 0 | 94 | 0 | 12 | 6 | 112 |
| 1997 | 0 | 0 | 83 | 0 | 11 | 12 | 106 |
| 1998 | 0 | 0 | 108 | 0 | 10 | 14 | 132 |
| 1999 | 0 | 0 | 126 | 0 | 13 | 18 | 157 |
| 2000 | 0 | 0 | 112 | 0 | 12 | 17 | 141 |
| 2001 | 0 | 0 | 73 | 0 | 13 | 12 | 98 |
| 2002 | 0 | 0 | 66 | 0 | 12 | 12 | 90 |
| 2003 | 0 | 0 | 99 | 1 | 12 | 16 | 128 |
| 2004 | 0 | 0 | 119 | 4 | 15 | 18 | 156 |
| 2005 | 0 | 0 | 101 | 3 | 16 | 13 | 133 |
| 2006 | 0 | 1 | 105 | 3 | 16 | 15 | 140 |
| 2007 | 0 | 1 | 119 | 3 | 15 | 20 | 158 |
| 2008 | 0 | 2 | 138 | 1 | 13 | 30 | 184 |
| 2009 | 0 | 1 | 98 | 1 | 14 | 33 | 147 |
| 2010 | 0 | 1 | 95 | 1 | 9 | 16 | 122 |
| 2011 | 0 | 1 | 103 | 0 | 15 | 12 | 131 |
| 2012 | 0 | 0 | 89 | 0 | 16 | 15 | 120 |
| 2013 | 0 | 0 | 70 | 0 | 9 | 13 | 92 |
| 2014 | 0 | 0 | 59 | 0 | 8 | 11 | 79 |
| 2015 | 0 | 0 | 104 | 11 | 8 | 19 | 143 |
|  |  |  |  |  |  |  |  |

Table 20.3 Landings ( $\mathbf{t}$ ) of brill Scophthalmus rhombus in Subdivisions 7.de (English Channel) by country, over the period 1950-2015 (Source: ICES Fishstat).

| Year | BEL | DNK | FRA | GBR | IRL | NLD | XCI | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 11 | 0 | 0 | 48 | 0 | 0 | 0 | 59 |
| 1951 | 8 | 0 | 0 | 70 | 0 | 0 | 0 | 78 |
| 1952 | 6 | 0 | 0 | 66 | 0 | 0 | 0 | 72 |
| 1953 | 2 | 0 | 0 | 60 | 0 | 0 | 0 | 62 |
| 1954 | 1 | 0 | 0 | 59 | 0 | 0 | 0 | 60 |
| 1955 | 4 | 0 | 0 | 57 | 0 | 0 | 0 | 61 |
| 1956 | 2 | 0 | 0 | 58 | 0 | 0 | 0 | 60 |
| 1957 | 4 | 0 | 0 | 66 | 0 | 0 | 0 | 70 |
| 1958 | 2 | 0 | 0 | 65 | 0 | 0 | 0 | 67 |
| 1959 | 1 | 0 | 0 | 58 | 0 | 0 | 0 | 59 |
| 1960 | 6 | 0 | 0 | 46 | 0 | 0 | 0 | 52 |
| 1961 | 1 | 0 | 0 | 46 | 0 | 0 | 0 | 47 |
| 1962 | 3 | 0 | 0 | 52 | 0 | 0 | 0 | 55 |
| 1963 | 1 | 0 | 0 | 50 | 0 | 0 | 0 | 51 |
| 1964 | 0 | 0 | 0 | 60 | 0 | 0 | 0 | 60 |
| 1965 | 2 | 0 | 0 | 46 | 0 | 0 | 0 | 48 |
| 1966 | 0 | 0 | 0 | 53 | 0 | 0 | 0 | 53 |
| 1967 | 1 | 0 | 0 | 66 | 0 | 0 | 0 | 67 |
| 1968 | 3 | 0 | 0 | 54 | 0 | 0 | 0 | 57 |
| 1969 | 2 | 0 | 121 | 67 | 0 | 0 | 0 | 190 |
| 1970 | 10 | 0 | 0 | 49 | 0 | 0 | 0 | 59 |
| 1971 | 18 | 0 | 0 | 48 | 0 | 0 | 0 | 66 |
| 1972 | 20 | 0 | 0 | 52 | 0 | 3 | 0 | 75 |
| 1973 | 20 | 0 | 0 | 70 | 0 | 0 | 0 | 90 |
| 1974 | 25 | 0 | 0 | 56 | 0 | 0 | 0 | 81 |
| 1975 | 24 | 0 | 55 | 56 | 0 | 0 | 2 | 137 |
| 1976 | 41 | 0 | 170 | 72 | 0 | 0 | 2 | 285 |
| 1977 | 45 | 0 | 197 | 77 | 0 | 0 | 4 | 323 |
| 1978 | 58 | 3 | 227 | 120 | 0 | 0 | 3 | 411 |
| 1979 | 55 | 0 | 262 | 140 | 0 | 0 | 2 | 459 |
| 1980 | 64 | 2 | 213 | 118 | 3 | 0 | 2 | 402 |
| 1981 | 83 | 0 | 271 | 130 | 0 | 0 | 6 | 490 |
| 1982 | 105 | 0 | 225 | 149 | 0 | 1 | 7 | 487 |
| 1983 | 107 | 0 | 234 | 181 | 0 | 1 | 3 | 526 |
| 1984 | 114 | 0 | 226 | 186 | 0 | 0 | 5 | 531 |
| 1985 | 94 | 0 | 213 | 177 | 0 | 0 | 10 | 494 |
| 1986 | 115 | 0 | 183 | 147 | 0 | 0 | 11 | 456 |
| 1987 | 126 | 0 | 216 | 141 | 0 | 0 | 10 | 493 |
| 1988 | 112 | 0 | 202 | 133 | 0 | 0 | 5 | 452 |
| 1989 | 89 | 0 | 213 | 121 | 0 | 0 | 2 | 425 |
| 1990 | 99 | 0 | 249 | 187 | 0 | 0 | 8 | 543 |
| 1991 | 81 | 0 | 249 | 140 | 0 | 0 | 0 | 470 |


| YEAR | BEL | DNK | FRA | GBR | IRL | NLD | XCI | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 82 | 0 | 223 | 151 | 0 | 0 | 7 | 463 |
| 1993 | 78 | 0 | 256 | 152 | 0 | 0 | 4 | 490 |
| 1994 | 88 | 0 | 227 | 170 | 0 | 0 | 5 | 490 |
| 1995 | 91 | 0 | 248 | 200 | 1 | 0 | 18 | 558 |
| 1996 | 105 | 0 | 240 | 253 | 0 | 0 | 10 | 608 |
| 1997 | 107 | 0 | 185 | 198 | 1 | 0 | 10 | 501 |
| 1998 | 70 | 0 | 196 | 173 | 0 | 2 | 10 | 451 |
| 1999 | 97 | 0 | 0 | 127 | 0 | 3 | 13 | 240 |
| 2000 | 164 | 0 | 260 | 232 | 1 | 4 | 17 | 678 |
| 2001 | 212 | 0 | 256 | 251 | 0 | 2 | 17 | 738 |
| 2002 | 204 | 0 | 268 | 227 | 0 | 1 | 16 | 716 |
| 2003 | 217 | 0 | 287 | 238 | 1 | 1 | 15 | 759 |
| 2004 | 165 | 0 | 259 | 223 | 1 | 3 | 15 | 666 |
| 2005 | 138 | 0 | 267 | 183 | 0 | 2 | 21 | 611 |
| 2006 | 180 | 0 | 281 | 170 | 0 | 3 | 15 | 649 |
| 2007 | 205 | 0 | 325 | 199 | 0 | 1 | 11 | 741 |
| 2008 | 154 | 0 | 225 | 199 | 0 | 2 | 13 | 593 |
| 2009 | 131 | 0 | 278 | 171 | 0 | 1 | 10 | 591 |
| 2010 | 145 | 0 | 340 | 198 | 0 | 1 | 11 | 695 |
| 2011 | 141 | 0 | 277 | 204 | 0 | 0 | 0 | 622 |
| 2012 | 121 | 0 | 263 | 232 | 0 | 1 | 0 | 617 |
| 2013 | 143 | 0 | 237 | 214 | 0 | 1 | 6 | 601 |
| 2014 | 165 | 0 | 243 | 232 | 0 | 1 | 10 | 651 |
| 2015 | 162 | 0 | 274 | 250 | 0 | 0 | 5 | 691 |

Table 20.4 Total international landings ( $\mathbf{t}$ ) of brill Scophthalmus rhombus in the Greater North Sea over the period 1950-2015, subdivided into Subarea 4 and Subdivisions 3.a and 7.d-e (Source: ICES Fishstat).

| Year |  |  |  | TOTAL |
| :---: | :---: | :---: | :---: | :---: |
| 1950 | 319 | 384 | 59 | 762 |
| 1951 | 337 | 511 | 78 | 926 |
| 1952 | 236 | 565 | 72 | 873 |
| 1953 | 246 | 589 | 62 | 897 |
| 1954 | 234 | 529 | 60 | 823 |
| 1955 | 212 | 571 | 61 | 844 |
| 1956 | 213 | 516 | 60 | 789 |
| 1957 | 148 | 468 | 70 | 686 |
| 1958 | 203 | 480 | 67 | 750 |
| 1959 | 233 | 424 | 59 | 716 |
| 1960 | 318 | 486 | 52 | 856 |
| 1961 | 305 | 581 | 47 | 933 |
| 1962 | 207 | 591 | 55 | 853 |
| 1963 | 120 | 620 | 51 | 791 |
| 1964 | 106 | 565 | 60 | 731 |
| 1965 | 155 | 535 | 48 | 738 |
| 1966 | 187 | 546 | 53 | 786 |
| 1967 | 106 | 409 | 67 | 582 |
| 1968 | 100 | 579 | 57 | 736 |
| 1969 | 99 | 658 | 190 | 947 |
| 1970 | 97 | 618 | 59 | 774 |
| 1971 | 104 | 1073 | 66 | 1243 |
| 1972 | 120 | 994 | 75 | 1189 |
| 1973 | 131 | 989 | 90 | 1210 |
| 1974 | 200 | 1152 | 81 | 1433 |
| 1975 | 187 | 1222 | 137 | 1546 |
| 1976 | 224 | 1208 | 285 | 1717 |
| 1977 | 388 | 1410 | 323 | 2121 |
| 1978 | 216 | 1375 | 411 | 2002 |
| 1979 | 184 | 1363 | 459 | 2006 |
| 1980 | 82 | 1033 | 402 | 1517 |
| 1981 | 59 | 1218 | 490 | 1767 |
| 1982 | 74 | 1294 | 487 | 1855 |
| 1983 | 83 | 1448 | 526 | 2057 |
| 1984 | 97 | 1522 | 531 | 2150 |
| 1985 | 110 | 1709 | 494 | 2313 |
| 1986 | 107 | 1207 | 456 | 1770 |
| 1987 | 105 | 970 | 493 | 1568 |
| 1988 | 101 | 1085 | 452 | 1638 |
| 1989 | 97 | 1302 | 425 | 1824 |
| 1990 | 128 | 893 | 543 | 1564 |
| 1991 | 99 | 1682 | 470 | 2251 |


|  | YEAR |  | 3.A |  | 4 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| TOTAL |  |  |  |  |  |
| 1992 | 146 | 1810 | 463 | 2419 |  |
| 1993 | 212 | 2439 | 490 | 3141 |  |
| 1994 | 222 | 1916 | 490 | 2628 |  |
| 1995 | 151 | 1434 | 558 | 2143 |  |
| 1996 | 112 | 1247 | 608 | 1967 |  |
| 1997 | 106 | 957 | 501 | 1564 |  |
| 1998 | 132 | 1283 | 451 | 1866 |  |
| 1999 | 157 | 1280 | 240 | 1677 |  |
| 2000 | 141 | 1508 | 678 | 2327 |  |
| 2001 | 98 | 1573 | 738 | 2409 |  |
| 2002 | 90 | 1302 | 716 | 2108 |  |
| 2003 | 128 | 1346 | 759 | 2233 |  |
| 2004 | 156 | 1249 | 666 | 2071 |  |
| 2005 | 133 | 1160 | 611 | 1904 |  |
| 2006 | 140 | 1175 | 649 | 1964 |  |
| 2007 | 158 | 1240 | 741 | 2139 |  |
| 2008 | 184 | 1004 | 593 | 1781 |  |
| 2009 | 147 | 1162 | 591 | 1900 |  |
| 2010 | 122 | 1500 | 695 | 2317 |  |
| 2011 | 131 | 1497 | 622 | 2250 |  |
| 2012 | 120 | 1512 | 617 | 2249 |  |
| 2013 | 92 | 1389 | 601 | 2082 |  |
| 2014 | 79 | 143 | 1655 | 651 | 1942 |
| 2015 | 143 |  | 691 | 2489 |  |

Table 20.5 Overall discard rates (all countries and métiers) for brill Scophthalmus rhombus in the Greater North Sea over the period 2012-2015 (Source: InterCatch).

| YEAR |  | DISCARD RATE |
| :--- | :--- | :--- |
| 2012 | 0,07 |  |
| 2013 | 0,04 |  |
| 2014 | 0,08 |  |
| 2015 | 0,09 |  |

Table 20.6 Discard rates for brill Scophthalmus rhombus in the Greater North Sea in 2015 by country (Source: InterCatch).

| Country |  | DISCARD RATE 2014 |  |
| :--- | :--- | :--- | :--- |
| Belgium | 0,01 | 0,03 | DISCARD RATE 2015 |
| Denmark | 0,21 | 0,22 | 0,24 |
| France | 0,04 | 0,01 |  |
| Germany | 0,16 | 0 |  |
| Ireland | 0 | 0,05 |  |
| Netherlands | 0,09 | 0 |  |
| Norway | 0 | 0,29 |  |
| Sweden | 0,35 | 0,02 |  |
| UK (England) | 0,01 | 0,2 |  |
| UK(Scotland) | 0,1 | 0,09 |  |
| Overall | 0,08 |  |  |

Table 20.7 Discard rates for brill Scophthalmus rhombus in the Greater North Sea in 2014 by Subarea/Division (Source: InterCatch).

|  | SUBAREA/ DIVISION | DISCARD RATE 2014 |  |
| :--- | :--- | :--- | :--- |
| 3.a | 0,38 | 0,33 |  |
| 4 | 0,08 | 0,04 |  |
| $7 . \mathrm{d}$ | 0,02 | 0,1 |  |
| $7 . \mathrm{e}$ | 0,01 | 0,14 |  |
| Overall | 0,08 | 0,09 |  |

Table 20.8 Survey index ( $\mathrm{N}^{\circ} / \mathrm{hr}$ ) for brill Scophthalmus rhombus in the BTS_ISI_Q3, Subarea 4.

|  | Year | N/hr | Year |
| :--- | :--- | :--- | :--- |
| 1987 | 1.9957265 | 2002 | 0.7947304 |
| 1988 | 0.6666667 | 2003 | 1.0000000 |
| 1989 | 0.9362745 | 2004 | 0.8214286 |
| 1990 | 2.2962963 | 2005 | 0.6060606 |
| 1991 | 1.8710526 | 2006 | 0.8716931 |
| 1992 | 3.6793860 | 2007 | 1.0952381 |
| 1993 | 3.3062753 | 2008 | 0.5138889 |
| 1994 | 2.3622590 | 2009 | 1.4246488 |
| 1995 | 1.8011775 | 2010 | 2.1853733 |
| 1996 | 0.7647059 | 2011 | 2.4057061 |
| 1997 | 2.0000000 | 2012 | 1.0411007 |
| 1998 | 1.4301503 | 2013 | 0.7586207 |
| 1999 | 0.7523810 | 2014 | 3.0445977 |
| 2000 | 2.1945342 | 2015 | 1.8429119 |
| 2001 | 0.6913580 |  |  |

Table 20.9 Survey index ( $\mathrm{N} / \mathrm{hr}$ ) for brill Scophthalmus rhombus in the BITS_HAF_Q1\&4, Division 3.a.

|  | Year |
| :--- | :--- |
| 1996 | 1.9090909 |
| 1997 | 0.3888889 |
| 1998 | 0.5000000 |
| 1999 | 1.8333333 |
| 2000 | 0.5555556 |
| 2001 | 1.0416667 |
| 2002 | 1.8030303 |
| 2003 | 1.3636364 |
| 2004 | 2.2045455 |
| 2005 | 2.0833333 |
| 2006 | 3.8181818 |
| 2007 | 3.6196970 |
| 2008 | 4.0500000 |
| 2009 | 3.0912698 |
| 2010 | 3.8893939 |
| 2011 | 3.6136364 |
| 2012 | 2.2651515 |
| 2013 | 2.1390227 |
| 2014 | 3.8551515 |
| 2015 | 4.4682540 |

Table 20.10 Commercial LPUE (kg/day) for brill Scophthalmus rhombus in the Dutch beam trawl fleet $>221 \mathrm{~kW}$, Subarea 4.

|  | YeAR | LPUE (KG/DAY) |
| :--- | :--- | :--- |
| 2007 | 33.38 |  |
| 2008 | 41.14 |  |
| 2009 | 40.65 |  |
| 2010 | 50.1 |  |
| 2011 | 52.39 |  |
| 2012 | 55.52 |  |
| 2013 | 52.97 |  |
| 2014 | 47.78 |  |
| 2015 | 60.74 |  |

## Brill Greater North Sea



Figure 20.1 Total international landings $(\mathbf{t})$ of brill Scophthalmus rhombus in the Greater North Sea over the period 1950-2015, subdivided into Subarea 4, Division 3.a and Divisions 7.de (Source: ICES Fishstat).


Figure 20.2 Relative contribution of landings of brill Scophthalmus rhombus from Subarea 4, Division 3.a and Divisions 7.de to the total international landings ( $\mathbf{t}$ ) in the Greater North Sea over the period 1950-2015 (Source: ICES Fishstat).


Fig. 20.3 Numbers of brill Scophthalmus rhombus caught per hour and rectangle by BTS_ISI_Q3 in the North Sea (4) over the period 1987-2015.


Fig. 20.4 Abundance index (numbers caught per hour) of brill Scophthalmus rhombus for the BTS_ISI_Q3 in the North Sea (4) over the period 1987-2015.


Figure 20.5 ALK of brill Scophthalmus rhombus in the North Sea (4) as documented in the BTS_ISI_Q3.


Figure 20.6 Length distributions (per 5 years) of brill Scophthalmus rhombus in the North Sea (4) as documented in the BTS_ISI_Q3.


Figure 20.7 Maturity at length of brill Scophthalmus rhombus in the North Sea (4) as documented in the BTS_ISI_Q3.


Fig. 20.8 Numbers of brill Scophthalmus rhombus caught per hour and rectangle by BITS_HAF_Q1\&4 in the Kattegat (3.aS) over the period 1996-2015.


Fig. 20.9 Abundance index (numbers caught per hour) of brill Scophthalmus rhombus for the BITS_HAF_Q1\&4 in the Kattegat (3.aS) over the period 1996-2015.


Figure 20.10 Length distributions (per 5 years) of brill Scophthalmus rhombus in the Kattegat (3.aS) as documented in the BITS_HAF_Q1\&4.


Fig. 20.11 Commercial LPUE (kg/day) of brill Scophthalmus rhombus in the Dutch beam trawl fleet $>221 \mathrm{~kW}$ (standardized for engine power and corrected for targeting behaviour). The bright red lines are the averages of the last two (2014-2015) and the previous three (2011-2013) years.

# 21 Dab (Limanda limanda) in Subarea 4 and Division 3.a (North Sea, Skagerrak and Kattegat) 

### 21.1 General

Dab (Limanda limanda) was assessed for the first time by the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK) in 2014. Until 2013 it was assessed by the Working Group on Assessment of New MoU Species (ICES, 2013.a). This group was dissolved in 2014. Because only official landings and survey data were available at that time, dab was defined as a category 3 species according to the ICES guidelines for data limited stocks (ICES, 2012). Since 2015 dab was included in the official data call for the WGNSSK and discard estimates were included into the dab assessment. Based on survey trends and total catch data (2012-2014) biennial advice for dab was given in 2015 (ICES, 2015). In 2016 a benchmark assessment of dab was conducted by ICES. For this benchmark assessment catch data from 2002 were requested and uploaded into the InterCatch data portal by all relevant countries (ICES, 2016). The benchmark agreed on the use of a survey based assessment model (SURBA) to inform stock status of North Sea dab (ICES, 2016). Based on these results the perception of the stock did not change and the WGNSSK agreed on not to reopen the advice for dab.

### 21.1.1 Biology and ecosystem aspects

Dab is a widespread demersal species on the Northeast Atlantic shelf and distributed from the Bay of Biscay to Iceland and Norway, including the Barents Sea and the Baltic. In the North Sea it is one of the most abundant species distributed over the whole area in depths down to 100 m , but it was also found occasionally down to depths of 150 m . The main concentration of dab can be found in the south eastern North Sea, especially the younger age groups 1-2. Older age groups are more distributed in the central and more Northern parts of the North Sea (Fig. 21.1.1.1). Dab abundance decreases towards the northern parts of the North Sea. Dab feeds on a variety of small invertebrates, mainly polychaete worms, shellfish and crustaceans. Early sexual maturation was reported for dab, maturing at ages of 2 to 3 years corresponding to approximately 11 to 14 cm total length. Peak spawning in the south eastern North Sea occurs from February to April.

### 21.1.2 Stock ID and possible assessment areas

The several spawning grounds and the wide distribution of dab indicate the presence of more than one stock. Meristic data (Lozán 1988) corroborate the hypothesis of several stocks for dab, distinguishing significantly between populations from western British waters, the North Sea and the Baltic Sea.

### 21.1.3 Management regulations

According to EU-Regulations a precautionary TAC is given in EU waters of 2a and 4 together with flounder (Plathichthys flesus). Since 2011 the TAC of 18434t did not change. No minimum landing size is defined. Dab is mainly a bycatch species in fisheries for plaice and sole. The discard rates for dab can be extremely high ( $\sim 90 \%$ ). TACs may not be appropriate as a management tool for bycatch species.

### 21.2 Fisheries data

### 21.2.1 Historical landings

Dab is a by-catch species in fisheries for plaice, sole and demersal round fish. According to ICES catch statistics, annual landings of dab in ICES Divisions 4 and 3.a has been well above 10000t since 1973 (Figure 21.2.1.1). The apparent decreases in official landings in the 1980's and 1990's are due to unreported catches by the Netherlands and Norway. However, since 1999 landings in area 4 and area 3.a steadily decreased. This trend continued in 2015 with total official landings of 4321 t .

The main fishing gear in the North Sea is the beam trawl with mesh sizes between 80 and 100 mm . Large effort reductions took place in this fishery over the last decade. The largest part of the landings in area 4 is taken by the Netherlands, followed by UK and Denmark (Figure 21.2.1.2). In division 3.a Denmark is landing the largest amount of dab (Figure 21.2.1.3). Dab is among the most discarded fish species in ICES Division 4. In the beam trawl fishery on plaice and sole and the otter trawl fishery on plaice up to $95 \%$ of the catches on dab are discarded (e.g. van Helmond et al. 2012).

### 21.2.2 InterCatch

For the WGNSSK2016 dab landing and discard data from 2002-2015 were available in the InterCatch web portal. Norway did not report any discards because of the official discard ban. Discard information in 2015 was provided for $56 \%$ of total landings in relation to weight calculated for the most disaggregated level of data (59\% in 2014, 59\% in 2013, $49 \%$ in 2012).

In 2015 the largest amount of landings and discards was again reported by The Netherlands for the TBB_DEF_70-99_0_0_all metier (Fig. 21.2.2.2 and Fig. 21.2.2.3). Consequently, by far the largest catch is taken by the Netherlands (27258t in total). All other countries catch less than 10000 t (Figure 21.2.2.4). The total dab catch estimated with InterCatch for 2015 was 52454 t from which 5082t were landings and 47372 t discards ( $90 \%$ of total catch). It should be noted that not all metiers were sampled in every quarter and that raising procedure may not be adequate in all cases. However, the Dutch TBB_DEF_70-99_0_0_all metier is by far the most important one in terms of landings and information on discard weights was provided for every quarter.

### 21.3 Survey data / recruit series

Surveys providing information on distribution, abundance and length frequency for dab in area 4 and division 3.a are the International Bottom Trawl Survey (IBTS) in quarter 1 and quarter 3 (Figure 21.3.1.), the Beam Trawl Surveys (BTS; only area 4) in quarter 3 and the BITS (only in division 3.a). To estimate a mature biomass index a length weight relationship derived from IBTS Q1 data was estimated in previous years to apply the DLS 3.2. method. The same data set was used to create a length based maturity ogive. The obtained length weight relationship and the maturity ogive were then applied to estimate the mature biomass index in kg per hour (Fig. 21.3.3.). From 1983 onwards the abundance index showed an increasing trend (Figure 21.4.2). Since the beginning of the 1990s the stock abundance index is fluctuating on a rather stable level. After a quite low index in 2012 the index was much higher for the last three years. This index served as an input for a survey based assessment model (SURBA) to inform the stock status of North Sea dab.

Only the beam trawl surveys provide data on age and weight for dab. During the benchmark in 2016 it was agreed to use an age based survey index combining data
from the Dutch and German beam trawl surveys taking into account a ship effect (Berg et al., 2014).

### 21.4 Survey Based Assessment (SURBA)

In spring 2016 a benchmark assessment was carried out for dab (ICES, 2016; see a summary in Annex 6). During this benchmark it was agreed to make use of the available data from the beam trawl surveys and to run a survey based assessment model taking the age structure into account. The final SURBAR (Needle, 2015) model run resulted in an overall decreasing total mortality while the spawning stock biomass (relative biomass) and the recruitment showed increasing trends (Fig. 21.4.1). The recruitment increased by a factor of 2.6 from 2003 to 2014 but dropped in 2015 . However, there is a strong retrospective pattern in recruitment with a general underestimation of recruitment for the terminal years (Fig. 21.4.4.). This might indicate a lower catchability of the survey for the youngest age group. No pattern was detected in the log residual pattern of the age based survey indices (Fig. 21.4.2.).

### 21.5 Analysis of stock trends

Dab is defined as a category 3 species following the ICES guidelines for data limited stocks (ICES, 2013). Consequently, the basis of the advice is a trend based assessment applying method 3.2 of the guidelines for data limited stocks:

$$
C_{y+1}=C_{y-1}\left(\frac{\sum_{i=y-x}^{y=y-1} I_{i} / x}{\sum_{i=y-z}^{y-x-1} I_{i} /(z-x)}\right)
$$

Where $C_{y+1}$ is the advised catch for the next year, $C_{y-1}$ should be the average catch of the last three years, and $I$ is the stock index. By default $x=2$ and $z=5$.

Table 21.4.2. displays the summary of the DLS 3.2. approach using the results of the updated and benchmarked assessment. However, the increasing trend in the SURBA SSB did not change the perception of this stock compared to the previous year. Therefore it was agreed not to reopen the advice for 2017.

### 21.6 References

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Lozán J.L., 1988. Verbreitung, Dichte, und Struktur der Population der Klieschen (Limanda limanda L.) in der Nordsee mit Vergleichen zu Popualtionen um Island und in der Ostsee anhand meristischer Merkmale. Arch. Fischereiwiss. 38: 165-189.

Table 21.4.1: Settings and input data used for the final SURBA assessment run.

| Setting/Data | Values/source |
| :--- | :--- |
| Survey index | Combined beam trawl survey index 2003 - current <br> assessment year (BTS-Isis, BTS-Tridens, German BTS) . Delta <br> GAM Method by Berg et al., 2014. |
| Ages | $1-6$ |
| Lambda | 3 |
| zbar | $1-6$ |
| Spawning time | 0.4 |
| Maturity ogive | Fixed ogive, age 1 = 60\%, age 2 = 80\%, age 3 and older 100\% |
| Weight at age | Data from Dutch Beam Trawl Surveys (2003 - current <br> assessment year) |

Table 21.4.1. Summary of the assessment.

| Year | Official LANDINGS | ICES <br> LANDINGS | ICES <br> CATCHES | ICES DISCARDS | IBTS Q1 <br> INDEX | DISCARD RATE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | 14771 |  |  |  | 12.15 |  |
| 1984 | 8251 |  |  |  | 11.96 |  |
| 1985 | 7047 |  |  |  | 13.04 |  |
| 1986 | 4813 |  |  |  | 18.02 |  |
| 1987 | 6189 |  |  |  | 22.52 |  |
| 1988 | 9321 |  |  |  | 22.09 |  |
| 1989 | 8162 |  |  |  | 29.97 |  |
| 1990 | 4275 |  |  |  | 32.28 |  |
| 1991 | 5057 |  |  |  | 20.86 |  |
| 1992 | 4101 |  |  |  | 30.91 |  |
| 1993 | 5004 |  |  |  | 32.95 |  |
| 1994 | 5822 |  |  |  | 22.35 |  |
| 1995 | 5395 |  |  |  | 28.31 |  |
| 1996 | 6239 |  |  |  | 20.97 |  |
| 1997 | 6271 |  |  |  | 18.92 |  |
| 1998 | 13720 |  |  |  | 21.61 |  |
| 1999 | 13949 |  |  |  | 19.56 |  |
| 2000 | 11249 |  |  |  | 16.18 |  |
| 2001 | 10564 |  |  |  | 16.32 |  |
| 2002 | 9655 | 8588 | 35219 | 26631 | 25.22 | 0.76 |
| 2003 | 9873 | 9433 | 54363 | 44930 | 26.04 | 0.83 |
| 2004 | 9387 | 8647 | 42920 | 34273 | 29.98 | 0.80 |
| 2005 | 10238 | 9537 | 44828 | 35291 | 23.51 | 0.79 |
| 2006 | 9914 | 10236 | 48214 | 37977 | 25.13 | 0.79 |
| 2007 | 10127 | 9881 | 43208 | 33328 | 33.09 | 0.77 |
| 2008 | 8551 | 8645 | 36024 | 27379 | 31.36 | 0.76 |
| 2009 | 7060 | 7040 | 40461 | 33421 | 22.81 | 0.83 |
| 2010 | 7830 | 8279 | 50765 | 42486 | 23.72 | 0.84 |
| 2011 | 7372 | 7422 | 51882 | 44460 | 25.32 | 0.86 |
| 2012 | 6749 | 7047 | 59679 | 52632 | 27.65 | 0.88 |
| 2013 | 6084 | 6611 | 60087 | 53476 | 20.28 | 0.89 |
| 2014 | 4957 | 5047 | 58780 | 53733 | 34.56 | 0.91 |
| 2015 | 4321 | 5082 | 52454 | 47372 | 33.59 | 0.90 |
| 2016 |  |  |  |  | 34.87 |  |

Table 21.4.2: Results of applying the DLS 3.2.

| INDICATOR (2011-2013) |  | 189.98 |
| :--- | :--- | :--- |
| Indicator (2014-2015) | 254.51 |  |
| Indicator ratio | 1.34 |  |
| Uncertainty cap | Yes |  |
| Average catch (2013-2015) | 57107 (tonnes) |  |
| Discard rate (2013-2015) | 0.9 |  |
| Precautionary buffer | No |  |
| Catch advice | No new advice for 2017 |  |

Table 21.6.1 Official dab landings by ICES area 4 and division 3.a.

| Year | 4 | $3 . \mathrm{A}$ | total |
| :---: | :---: | :---: | :---: |
| 1950 | 5971 | 1287 | 7258 |
| 1951 | 8190 | 1332 | 9522 |
| 1952 | 7976 | 1294 | 9270 |
| 1953 | 5915 | 1123 | 7038 |
| 1954 | 5652 | 1237 | 6889 |
| 1955 | 6623 | 1257 | 7880 |
| 1956 | 5468 | 2081 | 7549 |
| 1957 | 6127 | 2724 | 8851 |
| 1958 | 6342 | 2210 | 8552 |
| 1959 | 5239 | 1943 | 7182 |
| 1960 | 5168 | 1314 | 6482 |
| 1961 | 4602 | 1367 | 5969 |
| 1962 | 4082 | 1683 | 5765 |
| 1963 | 4615 | 1565 | 6180 |
| 1964 | 4982 | 1575 | 6557 |
| 1965 | 5519 | 2052 | 7571 |
| 1966 | 5862 | 1755 | 7617 |
| 1967 | 4324 | 1115 | 5439 |
| 1968 | 3995 | 1548 | 5543 |
| 1969 | 4122 | 1430 | 5552 |
| 1970 | 5183 | 1079 | 6262 |
| 1971 | 6546 | 1242 | 7788 |
| 1972 | 7901 | 1669 | 9570 |
| 1973 | 9657 | 1449 | 11106 |
| 1974 | 7146 | 2003 | 9149 |
| 1975 | 7033 | 2049 | 9082 |
| 1976 | 5917 | 1583 | 7500 |
| 1977 | 6702 | 2318 | 9020 |
| 1978 | 6407 | 2630 | 9037 |
| 1979 | 8243 | 2716 | 10959 |
| 1980 | 8357 | 2333 | 10690 |
| 1981 | 8454 | 2679 | 11133 |
| 1982 | 9565 | 2902 | 12467 |
| 1983 | 11865 | 2906 | 14771 |
| 1984 | 5482 | 2769 | 8251 |
| 1985 | 5502 | 1545 | 7047 |
| 1986 | 3205 | 1608 | 4813 |
| 1987 | 3931 | 2258 | 6189 |
| 1988 | 7067 | 2254 | 9321 |
| 1989 | 5816 | 2346 | 8162 |
| 1990 | 2701 | 1574 | 4275 |
| 1991 | 3448 | 1609 | 5057 |
| 1992 | 2647 | 1454 | 4101 |


|  | Year | 4 |  |
| :--- | :--- | :--- | :--- |

* preliminary catch statistics

Table 21.6.2 Official dab landings by country in area 4.

| Year | BEL | DEU | DNK | FRA | FRO | GBR | NLD | NOR | SWE | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 254 | 92 | 900 | 139 | 0 | 2555 | 2031 | 0 | 0 | 5971 |
| 1951 | 462 | 114 | 1800 | 90 | 0 | 3503 | 2221 | 0 | 0 | 8190 |
| 1952 | 386 | 74 | 1562 | 227 | 0 | 2823 | 2904 | 0 | 0 | 7976 |
| 1953 | 357 | 58 | 1337 | 189 | 0 | 2591 | 1383 | 0 | 0 | 5915 |
| 1954 | 255 | 62 | 1666 | 177 | 0 | 2393 | 1099 | 0 | 0 | 5652 |
| 1955 | 305 | 92 | 2923 | 161 | 0 | 1993 | 1149 | 0 | 0 | 6623 |
| 1956 | 338 | 99 | 1766 | 138 | 0 | 1660 | 1368 | 0 | 99 | 5468 |
| 1957 | 336 | 73 | 1983 | 154 | 0 | 1785 | 1669 | 0 | 127 | 6127 |
| 1958 | 290 | 71 | 2320 | 175 | 0 | 1885 | 1517 | 0 | 84 | 6342 |
| 1959 | 285 | 93 | 1433 | 146 | 0 | 2011 | 1265 | 0 | 6 | 5239 |
| 1960 | 246 | 70 | 1833 | 154 | 0 | 1813 | 1052 | 0 | 0 | 5168 |
| 1961 | 227 | 67 | 1497 | 161 | 0 | 1734 | 916 | 0 | 0 | 4602 |
| 1962 | 205 | 54 | 1357 | 147 | 0 | 1524 | 795 | 0 | 0 | 4082 |
| 1963 | 306 | 40 | 1660 | 128 | 0 | 1481 | 1000 | 0 | 0 | 4615 |
| 1964 | 424 | 48 | 1612 | 672 | 0 | 1177 | 1049 | 0 | 0 | 4982 |
| 1965 | 432 | 64 | 1841 | 734 | 0 | 1099 | 1349 | 0 | 0 | 5519 |
| 1966 | 507 | 65 | 1589 | 719 | 0 | 1215 | 1767 | 0 | 0 | 5862 |
| 1967 | 384 | 77 | 659 | 716 | 0 | 1147 | 1341 | 0 | 0 | 4324 |
| 1968 | 334 | 57 | 861 | 350 | 0 | 877 | 1516 | 0 | 0 | 3995 |
| 1969 | 302 | 69 | 984 | 448 | 0 | 689 | 1630 | 0 | 0 | 4122 |
| 1970 | 338 | 71 | 1476 | 588 | 0 | 752 | 1958 | 0 | 0 | 5183 |
| 1971 | 409 | 46 | 1546 | 618 | 0 | 986 | 2941 | 0 | 0 | 6546 |
| 1972 | 638 | 46 | 1816 | 727 | 0 | 1057 | 3617 | 0 | 0 | 7901 |
| 1973 | 678 | 41 | 1899 | 873 | 0 | 1349 | 3638 | 1179 | 0 | 9657 |
| 1974 | 281 | 59 | 1168 | 310 | 0 | 1227 | 4101 | 0 | 0 | 7146 |
| 1975 | 600 | 45 | 944 | 418 | 0 | 992 | 4031 | 0 | 3 | 7033 |
| 1976 | 489 | 52 | 852 | 306 | 0 | 816 | 3402 | 0 | 0 | 5917 |
| 1977 | 652 | 70 | 743 | 371 | 0 | 907 | 3959 | 0 | 0 | 6702 |
| 1978 | 520 | 64 | 799 | 513 | 0 | 1038 | 3473 | 0 | 0 | 6407 |
| 1979 | 484 | 87 | 1366 | 630 | 0 | 951 | 4724 | 0 | 1 | 8243 |
| 1980 | 518 | 24 | 1376 | 639 | 0 | 777 | 5023 | 0 | 0 | 8357 |
| 1981 | 542 | 31 | 1968 | 447 | 0 | 737 | 4729 | 0 | 0 | 8454 |
| 1982 | 460 | 42 | 2356 | 594 | 0 | 1002 | 5111 | 0 | 0 | 9565 |
| 1983 | 541 | 49 | 4428 | 495 | 0 | 1034 | 5318 | 0 | 0 | 11865 |
| 1984 | 603 | 35 | 3438 | 486 | 0 | 920 | 0 | 0 | 0 | 5482 |
| 1985 | 509 | 24 | 3535 | 404 | 0 | 1030 | 0 | 0 | 0 | 5502 |
| 1986 | 445 | 34 | 1400 | 289 | 0 | 1036 | 0 | 0 | 1 | 3205 |
| 1987 | 514 | 36 | 1574 | 434 | 0 | 1373 | 0 | 0 | 0 | 3931 |
| 1988 | 697 | 72 | 1324 | 349 | 0 | 1221 | 3404 | 0 | 0 | 7067 |
| 1989 | 443 | 117 | 1280 | 223 | 0 | 1232 | 2521 | 0 | 0 | 5816 |
| 1990 | 416 | 162 | 1103 | 214 | 0 | 802 | 0 | 0 | 4 | 2701 |
| 1991 | 491 | 290 | 1160 | 258 | 0 | 1249 | 0 | 0 | 0 | 3448 |
| 1992 | 464 | 218 | 699 | 217 | 0 | 1049 | 0 | 0 | 0 | 2647 |


| YEAR | BEL | DEU | DNK | FRA | FRO | GBR | NLD | NOR | SWE | 4 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1993 | 548 | 493 | 1016 | 235 | 0 | 1017 | 0 | 0 | 0 | 3309 |
| 1994 | 397 | 626 | 1307 | 133 | 0 | 1398 | 0 | 0 | 0 | 3861 |
| 1995 | 410 | 0 | 1306 | 155 | 1 | 1993 | 0 | 0 | 0 | 3865 |
| 1996 | 527 | 718 | 1484 | 177 | 0 | 1928 | 0 | 0 | 0 | 4834 |
| 1997 | 507 | 945 | 1399 | 124 | 0 | 2284 | 0 | 0 | 0 | 5259 |
| 1998 | 757 | 796 | 1024 | 126 | 0 | 2085 | 7971 | 0 | 0 | 12759 |
| 1999 | 802 | 758 | 1101 | 0 | 0 | 1964 | 8651 | 0 | 0 | 13276 |
| 2000 | 684 | 892 | 785 | 124 | 0 | 1534 | 6527 | 49 | 0 | 10595 |
| 2001 | 575 | 878 | 839 | 206 | 0 | 1368 | 5886 | 47 | 0 | 9799 |
| 2002 | 516 | 582 | 1126 | 228 | 0 | 1224 | 4951 | 51 | 0 | 8678 |
| 2003 | 396 | 642 | 1580 | 154 | 0 | 1204 | 4955 | 77 | 0 | 9008 |
| 2004 | 382 | 767 | 1136 | 121 | 0 | 1158 | 4989 | 55 | 0 | 8608 |
| 2005 | 372 | 1105 | 1128 | 121 | 0 | 1193 | 5352 | 131 | 0 | 9402 |
| 2006 | 369 | 1149 | 949 | 130 | 0 | 1415 | 5071 | 107 | 0 | 9190 |
| 2007 | 436 | 526 | 634 | 195 | 0 | 1212 | 6313 | 118 | 0 | 9434 |
| 2008 | 371 | 375 | 670 | 161 | 0 | 847 | 5544 | 61 | 0 | 8029 |
| 2009 | 349 | 262 | 489 | 196 | 0 | 648 | 4588 | 29 | 0 | 6561 |
| 2010 | 337 | 365 | 523 | 178 | 0 | 724 | 5097 | 16 | 0 | 7240 |
| 2011 | 243 | 312 | 622 | 165 | 0 | 645 | 4808 | 29 | 0 | 6824 |
| 2012 | 454 | 252 | 421 | 126 | 0 | 665 | 4136 | 41 | 0 | 6095 |
| 2013 | 404 | 333 | 404 | 84 | 0 | 647 | 3316 | 26 | 0 | 5214 |
| 2014 | 299 | 282 | 253 | 73 | 0 | 505 | 2910 | 23 | 0 | 4344 |
| $2015 *$ | 242 | 244 | 250 | 75 | 0 | 336 | 2438 | 10 | 0 | 3595 |
|  |  |  |  |  |  |  |  |  |  | 0 |

Table 21.6.3 Official dab landings in ICES division 3.a.

| YEAR | BEL | DEU | DNK | FRA | NLD | NOR | SWE | $3 . \mathrm{A}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0 | 34 | 1253 | 0 | 0 | 0 | 0 | 1287 |
| 1951 | 0 | 17 | 1315 | 0 | 0 | 0 | 0 | 1332 |
| 1952 | 0 | 21 | 1273 | 0 | 0 | 0 | 0 | 1294 |
| 1953 | 0 | 9 | 1114 | 0 | 0 | 0 | 0 | 1123 |
| 1954 | 0 | 4 | 1233 | 0 | 0 | 0 | 0 | 1237 |
| 1955 | 0 | 3 | 1254 | 0 | 0 | 0 | 0 | 1257 |
| 1956 | 0 | 5 | 1462 | 0 | 0 | 0 | 614 | 2081 |
| 1957 | 0 | 5 | 2025 | 0 | 0 | 0 | 694 | 2724 |
| 1958 | 0 | 4 | 1578 | 0 | 0 | 0 | 628 | 2210 |
| 1959 | 0 | 2 | 1307 | 0 | 0 | 0 | 634 | 1943 |
| 1960 | 0 | 1 | 1313 | 0 | 0 | 0 | 0 | 1314 |
| 1961 | 0 | 0 | 1367 | 0 | 0 | 0 | 0 | 1367 |
| 1962 | 0 | 2 | 1681 | 0 | 0 | 0 | 0 | 1683 |
| 1963 | 0 | 0 | 1565 | 0 | 0 | 0 | 0 | 1565 |
| 1964 | 0 | 1 | 1574 | 0 | 0 | 0 | 0 | 1575 |
| 1965 | 0 | 1 | 2051 | 0 | 0 | 0 | 0 | 2052 |
| 1966 | 0 | 0 | 1755 | 0 | 0 | 0 | 0 | 1755 |
| 1967 | 0 | 0 | 1115 | 0 | 0 | 0 | 0 | 1115 |
| 1968 | 0 | 0 | 1535 | 13 | 0 | 0 | 0 | 1548 |
| 1969 | 0 | 0 | 1430 | 0 | 0 | 0 | 0 | 1430 |
| 1970 | 0 | 0 | 1079 | 0 | 0 | 0 | 0 | 1079 |
| 1971 | 0 | 0 | 1242 | 0 | 0 | 0 | 0 | 1242 |
| 1972 | 0 | 0 | 1669 | 0 | 0 | 0 | 0 | 1669 |
| 1973 | 0 | 0 | 1449 | 0 | 0 | 0 | 0 | 1449 |
| 1974 | 0 | 0 | 2003 | 0 | 0 | 0 | 0 | 2003 |
| 1975 | 0 | 0 | 1959 | 0 | 2 | 0 | 88 | 2049 |
| 1976 | 10 | 0 | 1493 | 0 | 80 | 0 | 0 | 1583 |
| 1977 | 11 | 0 | 2105 | 0 | 142 | 0 | 60 | 2318 |
| 1978 | 2 | 0 | 2515 | 0 | 39 | 0 | 74 | 2630 |
| 1979 | 3 | 0 | 2616 | 0 | 15 | 0 | 82 | 2716 |
| 1980 | 3 | 0 | 2218 | 0 | 3 | 0 | 109 | 2333 |
| 1981 | 0 | 0 | 2574 | 0 | 5 | 0 | 100 | 2679 |
| 1982 | 1 | 0 | 2823 | 0 | 22 | 0 | 56 | 2902 |
| 1983 | 1 | 0 | 2759 | 0 | 34 | 0 | 112 | 2906 |
| 1984 | 0 | 0 | 2695 | 0 | 0 | 0 | 74 | 2769 |
| 1985 | 1 | 0 | 1486 | 0 | 0 | 0 | 58 | 1545 |
| 1986 | 5 | 0 | 1551 | 0 | 0 | 0 | 52 | 1608 |
| 1987 | 19 | 0 | 2182 | 0 | 0 | 0 | 57 | 2258 |
| 1988 | 13 | 0 | 2150 | 0 | 15 | 0 | 76 | 2254 |
| 1989 | 4 | 0 | 2302 | 0 | 0 | 0 | 40 | 2346 |
| 1990 | 3 | 0 | 1535 | 0 | 0 | 0 | 36 | 1574 |
| 1991 | 5 | 1 | 1556 | 0 | 0 | 0 | 47 | 1609 |
| 1992 | 10 | 0 | 1412 | 0 | 0 | 0 | 32 | 1454 |


| YEAR | BEL | DEU | DNK | FRA | NLD | NOR | SWE | 3.A |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1993 | 7 | 0 | 1656 | 0 | 0 | 0 | 32 | 1695 |
| 1994 | 9 | 0 | 1917 | 0 | 0 | 0 | 35 | 1961 |
| 1995 | 3 | 0 | 1482 | 0 | 0 | 0 | 45 | 1530 |
| 1996 | 0 | 0 | 1387 | 0 | 0 | 0 | 18 | 1405 |
| 1997 | 0 | 0 | 990 | 0 | 0 | 0 | 22 | 1012 |
| 1998 | 0 | 0 | 942 | 0 | 0 | 0 | 19 | 961 |
| 1999 | 0 | 0 | 661 | 0 | 0 | 0 | 12 | 673 |
| 2000 | 0 | 0 | 647 | 0 | 0 | 1 | 6 | 654 |
| 2001 | 0 | 0 | 751 | 0 | 0 | 7 | 7 | 765 |
| 2002 | 0 | 0 | 968 | 0 | 0 | 3 | 6 | 977 |
| 2003 | 0 | 0 | 674 | 0 | 173 | 14 | 4 | 865 |
| 2004 | 0 | 0 | 637 | 0 | 138 | 1 | 3 | 779 |
| 2005 | 0 | 0 | 738 | 0 | 95 | 0 | 3 | 836 |
| 2006 | 0 | 20 | 566 | 0 | 117 | 18 | 4 | 725 |
| 2007 | 0 | 9 | 547 | 0 | 126 | 3 | 9 | 694 |
| 2008 | 0 | 12 | 475 | 0 | 26 | 2 | 7 | 522 |
| 2009 | 0 | 4 | 478 | 0 | 3 | 1 | 12 | 498 |
| 2010 | 0 | 4 | 426 | 0 | 151 | 0 | 8 | 589 |
| 2011 | 0 | 10 | 517 | 0 | 0 | 11 | 7 | 545 |
| 2012 | 0 | 5 | 632 | 0 | 0 | 10 | 6 | 653 |
| 2013 | 0 | 11 | 654 | 0 | 174 | 26 | 6 | 871 |
| 2014 | 0 | 12 | 501 | 0 | 75 | 2 | 21 | 611 |
| $2015^{*}$ | 0.0 | 7.6 | 687.2 | 0.0 | 0.0 | 8.0 | 23.2 | 726 |
|  |  |  |  |  |  |  |  |  |



Figure 21.2.1.1 Total official landings of dab in area 4 and division 3.a in 1950-2015.


Figure 21.2.1.2 Official landings of dab in area 4 by country up to 2015.


Figure 21.2.1.3 Official landings of dab in division 3.a by country in 1950-2015.


Figure 21.2.2.1 InterCatch. Dab landings and discards (kg) provision for sub-area 4 and division 3.a by metier and country in 2015 as uploaded to InterCatch.


Figure 21.2.2.2 InterCatch. Dab landings (t) for sub-area 4 and division 3.a by metier and country in 2015 as uploaded to InterCatch.


Figure 21.2.2.3 InterCatch. Dab discards for sub-area 4 and division 3.a by metier and country in 2015. Reported discards (a), raised discards (b).


Figure 21.2.2.4 InterCatch. Dab landings and estimated discards for sub-area 4 and division 3.a by countries in 2015.


Fig. 21.3.1. Standardized dab survey indices (n/hour) from the beam trawl surveys.


Fig. 21.3.2. Standardized dab survey indices ( $n /$ hour) from the International Bottom Trawl Survey.


Fig. 21.3.3. Updated mature biomass index ( $\mathrm{kg} / \mathrm{h}$ ) as previously used for the DLS 3.2. method.


Fig. 21.3.4. Combined beam trawl index by age groups.


Fig. 21.3.3. Dab distribution in the North Sea by age group obtained by the Dutch and German Beam Trawl Surveys.


Fig. 21.4.1. SURBA model results for dab total mortality ( z ), spawning stock biomass (SSB), total stock biomass (TSB) and recruitment.


Fig. 21.4.2. SURBA model results of $\log$ residuals.


Fig. 21.4.3. SURBA model results displaying the age, year and cohort effects.


Fig. 21.4.4. SURBA model results. Retrospective runs.

## 22 Flounder (Platichtys flesus) in Subarea 4 and Division 3.a (North Sea, Skagerrak and Kattegat)

### 22.1 General

Flounder (Platichthys flesus) was assessed until 2013 in the Working Group on Assessment of New MoU Species (ICES, 2013.a). Because only official landings and survey data were available, flounder was defined as a category 3 species according to the ICES guidelines for data limited stocks (ICES, 2012). Biennial advice for flounder was given in 2013 by ICES (ICES, 2013b) based on survey trends. In 2014, flounder was included into the data call for the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK) and the biennial advice was evaluated in this group. The last biennial advice was that catches should be no more than 5228 tonnes in each of the years 2016 and 2017 (ICES, 2015). If discard rates do not change from the average of the last three years (2012-2014), this implies landings of no more than 2876 tonnes. The WGNSSK 2016 updated official landings, InterCatch raisings and the survey indices for flounder. The used survey index did not change the perception of the stock and thus it was concluded not to reopen the advice for 2017.

### 22.1.1 Biology and ecosystem aspects

Flounder is an euryhaline flatfish: the life cycle of each individual usually includes marine, brackish, and freshwater habitats. It has a coastal distribution in the Northeast Atlantic, ranging from the White Sea and the Baltic in the north, to the Mediterranean and Black Sea in the south. Flounder can live in low salinity water but they reproduce in water of higher salinity. In the North Sea, Skagerrak and Kattegat flounder spawn between February and April.
Flounder feeds on a wide variety of small invertebrates (mainly polychaete worms, shellfish, and crustaceans), but locally the diet may include small fish species like smelt and gobies. The most intensive feeding occurs in the summer, while food is sparse in the winter.

During autumn, both mature and immature flounder withdraw from the inshore and estuarine feeding areas. Immature flounder migrate into coastal areas, where they spend the winter. The adults move further offshore to the $25-40 \mathrm{~m}$ deep spawning grounds, the most important of which are situated along the coasts of Belgium, the Netherlands, Germany, and Denmark.

More details on available data and knowledge can be found in the flounder stock annex.

### 22.1.2 Stock ID and possible assessment areas

There is no information about stock identity and possible stock assessment areas in the North Sea, Skagerrak and Kattegat. Within the North Sea there may exist a number of sub-populations (WGNEW, 2013).

### 22.1.3 Management regulations

There is no minimum landing size for this species in EC waters.
In the EC waters of area 2 a and 4 there is a combined TAC for flounder and dab. The TAC for both species of 18.434 t was not changed in the last three years.

### 22.2 Fisheries data

### 22.2.1 Historical landings

In the North Sea and in Skagerrak-Kattegat flounder is mainly a by-catch in the fishery for commercially more important flatfish such as sole and plaice and in the mixed demersal fisheries. The largest part of official landings is reported for Subarea 4, especially for the last decade (Fig. 22.2.1.1; Table 22.2.1.3). Landings in ICES Subarea 4 and Division 3.a by country are shown in Figures 22.2.1.2 and 22.2.1.3 and in Tables 22.2.1.1 and 22.2.1.2. From Figure 22.2.1.1 it can be seen that the landings data are not complete: there is a gap in Dutch landings data for the time period 1984 to 1997.

Since 1950, annual landings from the North Sea have fluctuated, without a clear pattern (Figure 22.2.1.1). In the last decade, landings declined considerably. This decline goes hand in hand with a reduction in fishing effort of bottom trawl fleets in the North Sea. For 2015, total official landings were reported with 1883 t , compared to 2062 t in 2014. In area 3.a, annual landings have decreased sharply from 194 t in 2014 to 77 t in 2015 (Figure 22.2.1.3). In the beginning of the time series the landings seem to be fluctuating without a clear trend, however in last two decades the trend is declining. Flounder is of relatively little commercial importance in the North Sea and the Skagerrak-Kattegat. In the North Sea and the Skagerrak-Kattegat the landings data may be misreported in years that quota for commercially more important species are limited. The amount of misreporting however is not known. In addition, the North Sea landings may not reflect the catches very well. Flounder is often discarded and discarding is influenced by the prices and the availability of other, commercially more important species.

### 22.2.2 Inter Catch

In 2014 flounder was included for the first time into the data call for WGNSSK 2014. In 2016 data to cover the years 2012 - 2015 are available in InterCatch. From all countries data were uploaded to the Inter Catch data portal. Norway, France, and Scotland did only report landings but no discards for flounder. For the year 2012 The Netherlands provided only discard data. In general it was tried only to use equivalent or similar metiers for the raising procedure. Discard information was provided for $88 \%$ in relation to weight of total landings in 2015 ( $90 \%$ in 2014; 90\% in 2013).

In 2015 by far the largest proportion of landings ( $1365 \mathrm{t}, \sim 72 \%$ of total landings) was reported by the Netherlands and Belgium beam trawlers (TBB_DEF_70_99_0_0_all). Other metiers landing flounder in considerable amounts did not land more than 100t. These metiers were also dominated by Dutch landings (Fig. 22.2.2.1). The highest amount of discard in 2015 comes again from the MIS_MIS_0_0_0_HC metier (Fig. 22.2.2.2.).

A problem in the estimation of total flounder discards maybe the TBB_CRU_1632_0_0_all metier targeting brown shrimps in more coastal areas. For this metier relatively high discards but extremely low landings were reported by Germany. The Netherlands and Belgium reported landings but no discards. It was not meaningful to use the German fleet to raise the Belgium and Dutch landings which would probably have resulted in unrealistic high discards for these fleets. However, given the amount discarded by Germany and the similar effort in this metier by The Netherlands this might lead to a substantial underestimation of the total discard estimation. It might be useful in the future to raise discard by effort for these fleets and also for some metiers with zero landings for which no discards can be raised although they might occur in these metiers.

The highest total catch is taken by the Netherlands, followed by Scotland (nearly all reported discards for Scotland). Belgium and Denmark also take a considerable part of the catch while all other countries catch less than 100t (Fig. 22.2.2.3). The total catch estimated with Inter Catch was 3045 t from which 1762 t were landings (compared to 1883 t reported official landings) and 1283 t discards ( $42 \%$ of total catches which is the same value as for the last year). However, it should be noted that not all metiers were sampled in every quarter and that the raising procedure may not be adequate in all cases.

### 22.3 Survey data / recruit series

Several surveys in the North Sea, Skagerrak and Kattegat provide information on distribution, abundance and length composition of flounder. The most relevant survey for flounder is probably the International Bottom Trawl Survey IBTS in quarter 1 (Figure 22.3.1 and Figure 22.4.3). However, the IBTS Q1 uses a bottom trawl which is not very well suited to catch demersal flatfishes. The BTS surveys use a beam trawl, but they are carried out in quarter 3, in a time of year in which flounder is usually distributed in more coastal, shallow and brackish waters. Therefore, it was decided by WGNEW2013 to use the IBTS Q1 to analyse survey trends for this species. It should be noted here that for the IBTS the gear in use was fully standardized since 1983. Therefore, index data before this year should be interpreted with caution and are not presented in this report.

### 22.4 Analysis of stock trends / assessment

In 2013 flounder was assessed in the Working Group on Assessment of New MoU Species (ICES, 2013). Until then, only landings data and survey trends were available for this species. Therefore, flounder was defined as a category 3 species following the ICES guidelines for data limited stocks (ICES, 2012). Consequently, the basis of the advice was a trend based assessment applying method 3.2 of the guidelines for data limited stocks:

$$
C_{y+1}=C_{y-1}\left(\frac{\sum_{i=y-x}^{y=y-1} I_{i} / x}{\sum_{i=y-z}^{y-x-1} I_{i} /(z-x)}\right)
$$

Where $C_{y+1}$ is the advised catch for the next year, $C_{y-1}$ should be the average catch of the last three years, and $I$ is the stock index. By default $x=2$ and $z=5$. A mature biomass index in kg per hour was estimated from the IBTS Q1 survey (excluding round fish areas 1 and 2), because this survey covers most of the distribution area of flounder in area 4 and 3.a.

As the used index for flounder did not change the perception of the stock it was agreed by the WGNSSK2016 that the advice was not reopened.

## Estimated indices

The mature biomass index ( $\mathrm{kg} / \mathrm{hour}$ ) estimated by WGNSSK2016 was based on the IBTS Q1 survey which covers most of the distribution area of flounder in area 4 and 3.a. Roundfish areas 1 and 2 were excluded from analyses because flounder does only occur very occasionally in these areas (Fig. 22.3.1). To estimate a mature biomass index (kg/hour) a length weight relationship derived from IBTS Q1 data was applied. The same data set shows that above 20 cm probably most flounder are mature (Fig. 22.4.1).

The biomass index shows a rather stable trend from 1983 onwards with two major peaks between 1985 and 1995 (Fig. 22.4.3). From 1996 to 2001 the index declined, followed by an increase until 2005. Since then it fluctuated without a clear trend up to 2010. A declining trend can be observed from 2010 to 2014, while the values of 2015 and 2016 are again somewhat higher.

### 22.5 References

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ICES 2013.a. Report of the Working Group on Assessment of New MoU Species (WGNEW), 2428 March 2013, ICES Headquarters, Denmark. ICES CM 2013/ACOM 21.

ICES 2013b. Flounder in Subarea IV and Division 3.a, Report of the ICES Advisory Committee, 2013. ICES Advice, 2013. Book 6, Section 6.4.29.

ICES. 2013c. Report of the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), 24-30 April 2013. ICES CM 2013/ACOM:13.

ICES 2012. ICES implementation of advice for data limited stocks in 2012. Report in support of ICES advice. ICES CM2012/ACOM:68.

Table 22.2.1.1 Flounder official landings by country in ICES area 4.

| Year | Belgium | Denmark | France | Germany | Netherlands | Norway | UK | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 67 | 1514 | 0 | 641 | 937 | 0 | 67 | 241 | 3467 |
| 1951 | 119 | 1143 | 0 | 329 | 949 | 0 | 81 | 127 | 2748 |
| 1952 | 91 | 1210 | 0 | 257 | 841 | 0 | 71 | 186 | 2656 |
| 1953 | 270 | 1372 | 0 | 397 | 886 | 0 | 92 | 203 | 3220 |
| 1954 | 142 | 1225 | 0 | 281 | 696 | 0 | 71 | 121 | 2536 |
| 1955 | 145 | 1244 | 0 | 353 | 871 | 0 | 88 | 109 | 2810 |
| 1956 | 132 | 1389 | 0 | 277 | 1097 | 0 | 102 | 2 | 2999 |
| 1957 | 81 | 910 | 0 | 250 | 825 | 0 | 112 | 0 | 2178 |
| 1958 | 99 | 784 | 0 | 257 | 1088 | 0 | 94 | 0 | 2322 |
| 1959 | 62 | 533 | 0 | 424 | 857 | 0 | 79 | 1 | 1956 |
| 1960 | 82 | 614 | 0 | 540 | 733 | 0 | 49 | 8 | 2026 |
| 1961 | 68 | 776 | 0 | 390 | 579 | 0 | 81 | 13 | 1907 |
| 1962 | 37 | 1146 | 0 | 313 | 717 | 0 | 53 | 2 | 2268 |
| 1963 | 16 | 501 | 0 | 263 | 467 | 0 | 65 | 0 | 1312 |
| 1964 | 30 | 1141 | 0 | 305 | 563 | 0 | 48 | 6 | 2093 |
| 1965 | 121 | 1349 | 0 | 248 | 549 | 0 | 54 | 3 | 2324 |
| 1966 | 32 | 946 | 0 | 229 | 573 | 0 | 71 | 2 | 1853 |
| 1967 | 43 | 540 | 0 | 193 | 331 | 0 | 57 | 25 | 1189 |
| 1968 | 75 | 894 | 0 | 152 | 160 | 0 | 43 | 1 | 1325 |
| 1969 | 54 | 582 | 0 | 158 | 161 | 0 | 33 | 0 | 988 |
| 1970 | 50 | 316 | 0 | 135 | 405 | 0 | 57 | 0 | 963 |
| 1971 | 60 | 685 | 0 | 173 | 297 | 0 | 70 | 0 | 1285 |
| 1972 | 63 | 991 | 0 | 159 | 275 | 0 | 60 | 0 | 1548 |
| 1973 | 63 | 290 | 0 | 172 | 1424 | 0 | 53 | 0 | 2002 |
| 1974 | 115 | 766 | 0 | 190 | 2661 | 0 | 58 | 0 | 3790 |
| 1975 | 68 | 437 | 0 | 155 | 2191 | 0 | 87 | 1 | 2939 |
| 1976 | 94 | 575 | 0 | 209 | 2077 | 0 | 70 | 54 | 3079 |
| 1977 | 107 | 320 | 0 | 208 | 1732 | 0 | 127 | 11 | 2505 |
| 1978 | 122 | 203 | 0 | 198 | 1519 | 0 | 169 | 0 | 2211 |
| 1979 | 129 | 181 | 31 | 275 | 1260 | 0 | 201 | 0 | 2077 |
| 1980 | 190 | 300 | 33 | 229 | 806 | 0 | 140 | 0 | 1698 |
| 1981 | 164 | 669 | 14 | 200 | 1068 | 0 | 133 | 0 | 2248 |
| 1982 | 110 | 630 | 31 | 200 | 1597 | 0 | 121 | 0 | 2689 |
| 1983 | 88 | 564 | 36 | 197 | 2059 | 0 | 125 | 0 | 3069 |
| 1984 | 272 | 518 | 15 | 103 | 0 | 0 | 122 | 0 | 1030 |
| 1985 | 163 | 379 | 14 | 128 | 0 | 0 | 109 | 0 | 793 |
| 1986 | 155 | 456 | 1 | 91 | 0 | 0 | 111 | 0 | 814 |
| 1987 | 132 | 394 | 32 | 106 | 0 | 0 | 90 | 0 | 754 |
| 1988 | 160 | 509 | 44 | 105 | 682 | 0 | 98 | 0 | 1598 |
| 1989 | 200 | 632 | 28 | 95 | 916 | 0 | 80 | 0 | 1951 |
| 1990 | 153 | 467 | 69 | 147 | 0 | 0 | 45 | 0 | 881 |
| 1991 | 260 | 377 | 51 | 902 | 0 | 0 | 69 | 0 | 1659 |
| 1992 | 152 | 492 | 35 | 521 | 0 | 0 | 76 | 0 | 1276 |


| Year | Belgium | Denmark | France | Germany | Netherlands | Norway | UK | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 194 | 1812 | 47 | 356 | 0 | 0 | 136 | 0 | 2545 |
| 1994 | 196 | 642 | 57 | 921 | 0 | 0 | 247 | 0 | 2063 |
| 1995 | 301 | 628 | 103 | 843 | 0 | 0 | 250 | 0 | 2125 |
| 1996 | 262 | 1439 | 68 | 43 | 0 | 0 | 193 | 0 | 2005 |
| 1997 | 110 | 988 | 10 | 25 | 0 | 0 | 157 | 0 | 1290 |
| 1998 | 283 | 154 | 40 | 13 | 4938 | 0 | 132 | 0 | 5560 |
| 1999 | 326 | 123 | 0 | 11 | 3158 | 0 | 54 | 0 | 3672 |
| 2000 | 289 | 100 | 46 | 17 | 2656 | 5 | 52 | 0 | 3165 |
| 2001 | 241 | 92 | 42 | 4 | 2608 | 3 | 32 | 0 | 3022 |
| 2002 | 165 | 83 | 51 | 2 | 3531 | 3 | 55 | 0 | 3890 |
| 2003 | 206 | 94 | 33 | 3 | 3172 | 9 | 120 | 0 | 3637 |
| 2004 | 335 | 96 | 46 | 5 | 3720 | 18 | 74 | 0 | 4294 |
| 2005 | 241 | 171 | 17 | 5 | 3363 | 38 | 111 | 0 | 3946 |
| 2006 | 168 | 152 | 19 | 2 | 4020 | 39 | 216 | 0 | 4616 |
| 2007 | 298 | 166 | 56 | 45 | 2925 | 11 | 119 | 0 | 3620 |
| 2008 | 306 | 228 | 30 | 39 | 2231 | 3 | 57 | 0 | 2894 |
| 2009 | 272 | 273 | 38 | 46 | 2124 | 3 | 59 | 0 | 2815 |
| 2010 | 251 | 126 | 20 | 58 | 2612 | 6 | 87 | 0 | 3160 |
| 2011 | 262 | 112 | 17 | 25 | 2566 | 1 | 65 | 0 | 3048 |
| 2012 | 348 | 100 | 11 | 23 | 1672 | 0 | 38 | 0 | 2192 |
| 2013 | 346 | 93 | 13 | 28 | 1199 | 0 | 24 | 0 | 1703 |
| 2014 | 366 | 107 | 15 | 30 | 1318 | 1 | 31 | 0 | 1868 |
| 2015* | 301 | 97 | 18 | 19 | 1356 | 15 | 0 | 0 | 1806 |

*Preliminary catch statistics

Table 22.2.1.2 Flounder official landings by country in ICES division 3.a.

| Year | Denmark | Germany | Netherlands | Norway | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 1632 | 92 | 0 | 0 | 657 | 2381 |
| 1951 | 1548 | 88 | 0 | 0 | 759 | 2395 |
| 1952 | 1161 | 48 | 0 | 0 | 683 | 1892 |
| 1953 | 1135 | 17 | 0 | 0 | 724 | 1876 |
| 1954 | 1138 | 13 | 0 | 0 | 528 | 1679 |
| 1955 | 1265 | 11 | 0 | 0 | 667 | 1943 |
| 1956 | 1229 | 6 | 0 | 0 | 0 | 1235 |
| 1957 | 1331 | 12 | 0 | 0 | 0 | 1343 |
| 1958 | 1099 | 12 | 0 | 0 | 0 | 1111 |
| 1959 | 1003 | 3 | 0 | 0 | 0 | 1006 |
| 1960 | 875 | 10 | 0 | 0 | 566 | 1451 |
| 1961 | 821 | 9 | 0 | 0 | 442 | 1272 |
| 1962 | 812 | 3 | 0 | 0 | 0 | 815 |
| 1963 | 554 | 0 | 0 | 0 | 0 | 554 |
| 1964 | 822 | 1 | 0 | 0 | 0 | 823 |
| 1965 | 1016 | 0 | 0 | 0 | 0 | 1016 |
| 1966 | 1027 | 0 | 0 | 0 | 0 | 1027 |
| 1967 | 811 | 3 | 0 | 0 | 0 | 814 |
| 1968 | 808 | 2 | 0 | 0 | 0 | 810 |
| 1969 | 721 | 0 | 0 | 0 | 0 | 721 |
| 1970 | 667 | 0 | 0 | 0 | 0 | 667 |
| 1971 | 611 | 1 | 0 | 0 | 0 | 612 |
| 1972 | 365 | 0 | 0 | 0 | 0 | 365 |
| 1973 | 346 | 0 | 0 | 0 | 0 | 346 |
| 1974 | 1656 | 2 | 0 | 0 | 0 | 1658 |
| 1975 | 1377 | 1 | 0 | 0 | 89 | 1467 |
| 1976 | 949 | 2 | 4 | 0 | 144 | 1099 |
| 1977 | 1036 | 0 | 19 | 0 | 64 | 1119 |
| 1978 | 1560 | 10 | 14 | 0 | 64 | 1648 |
| 1979 | 1219 | 0 | 0 | 0 | 100 | 1319 |
| 1980 | 426 | 0 | 0 | 0 | 135 | 561 |
| 1981 | 1831 | 0 | 0 | 0 | 74 | 1905 |
| 1982 | 1236 | 0 | 0 | 0 | 75 | 1311 |
| 1983 | 2352 | 0 | 0 | 0 | 160 | 2512 |
| 1984 | 2463 | 0 | 0 | 0 | 283 | 2746 |
| 1985 | 1203 | 0 | 0 | 0 | 102 | 1305 |
| 1986 | 1585 | 0 | 0 | 0 | 166 | 1751 |
| 1987 | 1050 | 0 | 0 | 0 | 119 | 1169 |
| 1988 | 1164 | 0 | 0 | 0 | 149 | 1313 |
| 1989 | 996 | 0 | 0 | 0 | 133 | 1129 |
| 1990 | 650 | 1 | 0 | 0 | 57 | 708 |
| 1991 | 574 | 0 | 0 | 0 | 50 | 624 |
| 1992 | 455 | 0 | 0 | 0 | 52 | 507 |


| Year | Denmark | Germany | Netherlands | Norway | SWeden | Total |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- |
| 1993 | 673 | 3 | 0 | 0 | 67 | 743 |
| 1994 | 865 | 1 | 0 | 0 | 77 | 943 |
| 1995 | 403 | 19 | 0 | 0 | 76 | 498 |
| 1996 | 429 | 9 | 0 | 0 | 104 | 542 |
| 1997 | 367 | 2 | 0 | 0 | 68 | 437 |
| 1998 | 637 | 5 | 0 | 0 | 83 | 725 |
| 1999 | 558 | 6 | 0 | 0 | 24 | 588 |
| 2000 | 609 | 17 | 0 | 0 | 30 | 656 |
| 2001 | 672 | 2 | 0 | 1 | 30 | 705 |
| 2002 | 493 | 0 | 0 | 1 | 30 | 524 |
| 2003 | 452 | 3 | 0 | 0 | 18 | 473 |
| 2004 | 462 | 2 | 0 | 0 | 14 | 478 |
| 2005 | 467 | 0 | 0 | 0 | 15 | 482 |
| 2006 | 380 | 0 | 0 | 0 | 13 | 393 |
| 2007 | 419 | 3 | 1 | 0 | 22 | 445 |
| 2008 | 326 | 4 | 0 | 0 | 16 | 346 |
| 2009 | 238 | 2 | 0 | 0 | 33 | 273 |
| 2010 | 188 | 0 | 0 | 0 | 17 | 205 |
| 2011 | 129 | 0 | 0 | 0 | 16 | 145 |
| 2012 | 110 | 0 | 0 | 0 | 8 | 118 |
| 2013 | 162 | 0 | 0 | 0 | 11 | 173 |
| 2014 | 190 | 0 | 0 | 0 | 4 | 194 |
| $2015^{*}$ | 74 | 0 | 0 | 0 | 3 | 77 |
|  |  |  |  |  |  |  |

*preliminary catch statistics

Table 22.2.1.3 Flounder total official landings by ICES areas.


| Year |  |  | 4 | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1993 | 743 | 2545 |  | 3288 |
| 1994 | 943 | 2063 |  | 3006 |
| 1995 | 498 | 2125 |  | 2623 |
| 1996 | 542 | 2005 |  | 2547 |
| 1997 | 437 | 1290 |  | 1727 |
| 1998 | 725 | 5560 |  | 6285 |
| 1999 | 588 | 3672 |  | 4260 |
| 2000 | 656 | 3165 |  | 3821 |
| 2001 | 705 | 3022 |  | 3727 |
| 2002 | 524 | 3890 |  | 4414 |
| 2003 | 473 | 3637 |  | 4110 |
| 2004 | 478 | 4294 |  | 4772 |
| 2005 | 482 | 3946 |  | 4428 |
| 2006 | 393 | 4616 |  | 5009 |
| 2007 | 445 | 3620 |  | 4065 |
| 2008 | 346 | 2894 |  | 3240 |
| 2009 | 273 | 2815 |  | 3088 |
| 2010 | 205 | 3160 |  | 3365 |
| 2011 | 145 | 3048 |  | 3193 |
| 2012 | 118 | 2192 |  | 2310 |
| 2013 | 173 | 1703 |  | 1876 |
| 2014 | 194 | 1868 |  | 2062 |
| 2015* | 77 | 1806 |  | 1883 |

*preliminary catch statistics


Fig. 22.2.1.1 Official landings of flounder by area.


Fig. 22.2.1.2 Official landings of flounder in ICES area 4 by country.


Fig. 22.2.1.3 Official landings of flounder in ICES division 3.a by country.


Fig. 22.2.2.1 Inter Catch. Flounder landings by metier and country in 2015 as uploaded to Inter Catch.


Fig. 22.2.2.2 Inter Catch. Flounder discards by metier and country in 2015. Reported discards panel (a), raised discards panel (b).


Fig. 22.2.2.3 Inter Catch. Flounder landings and discards by country in 2014.


Fig. 22.3.1 Distribution of flounder derived from the IBTS Q1 survey in area 4 and division 3.a for the whole time series.


Fig. 22.4.1 Length weight relationship of flounder derived from IBTS Q1 data.


Fig. 22.4.2 Maturity at length of female and male flounder derived from IBTS Q1 data.

Flounder IBTS Q1 index


Fig. 22.4.3 Mature biomass index of flounder in area 4 and division 3.a derived from IBTS Q1 data 1983-2016.

## 23 Lemon sole (Microstomus kitt) in Subarea 4 and Divisions 3.a and 7.d (North Sea, Skagerrak and Kattegat, Eastern English Channel)

### 23.1 General

Until 2014, lemon sole (Microstomus kitt) was assessed in the Working Group on Assessment of New MoU Species (ICES, 2013). Lemon sole has been defined as a category 3 species according to the ICES guidelines for data limited stocks (ICES, 2012). Biennial advice for lemon sole was given in 2013 (ICES, 2013b), based on survey trends. In 2014, lemon sole was included to the data call for the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK). This is the third year in which the stock status for lemon sole has been evaluated by WGNSSK.

### 23.1.1 Biology and ecosystem aspects

Lemon sole is a commercially important flatfish that is found in the shelf waters of the North Atlantic from the White Sea and Iceland southwards to the Bay of Biscay. In Scottish waters, lemon sole spawn in the northwest of the North Sea in April and spawning spreads north and east as the season progresses (Rae, 1965). In the western English Channel, lemon sole spawn in April and May (Jennings et al. 1993). In the English Channel, investigations of habitat association for plaice, sole and lemon sole indicated that distribution is restricted to a few sites and that lemon soles appear to prefer sandy and gravely strata, living deeper and at a higher salinity and lower temperature than plaice or sole (Hinz et al., 2006). Lemon sole feeds on small invertebrates, mainly polychaete worms, bivalves and crustaceans.

### 23.1.2 Stock ID and possible assessment areas

There is little information available on lemon sole stock identity for the greater North Sea.

### 23.1.3 Management regulations

No specific management objectives are known to ICES. An EU TAC is set for EU waters of ICES Division 2.a and Subarea 4 together with witch flounder (ICES, 2013).

### 23.2 Fisheries data

### 23.2.1 Historical landings

In the North Sea and in Skagerrak-Kattegat lemon sole is mainly a by-catch species in the fishery for plaice and in the mixed demersal fisheries. Landings in ICES Division 7.d, and sub-area 4 and division 3.a are shown in Figures 23.2.1 to 23.2.4, and in Tables 23.2.1 to 23.2.4. The time-series of landings are not fully complete, and a number of countries have gaps in data provision.

### 23.2.2 Discards

Catch yields and age compositions for lemon sole for 2013 and 2014 were submitted for the first time to the InterCatch system prior to the 2015 WGNSSK meeting, enabling the estimation of discard rates (by weight) for those years. However, no age-sampled landings or discards were submitted for the 2016 WGNSSK meeting. Previously, only
around $10 \%$ of metiers landing lemon sole were age-sampled for landings and discards, and all such sampling was carried out by England. In 2016, age samples were not available. However, the majority of landings have estimates of total discards by weight associated with them. The time-series of official landings for the full stock area, along with WG estimates of landings and discards for 2013, 2014 and 2015, are given in Figure 23.2.5. Discard rates by weight in 2015 (29\%) remain similar to that of 2013 ( $28 \%$ ) and 2014 (31\%).

### 23.3 Survey data / recruit series

Surveys providing information on distribution, abundance and length frequency for lemon sole in area 4, division 7.d and division 3.a are the International Bottom Trawl Survey (IBTS) in quarter 1 and quarter 3, and the Beam Trawl Survey (BTS; only area 4) in quarter 3. The IBTS Q1 was used in the case of lemon sole to analyse stock trends. This survey uses a GOV demersal trawl which may not be the optimal gear to catch flatfish such as lemon sole. However, the beam trawl surveys do not cover the whole distribution area of lemon sole (missing the northern area in particular, see Figure 23.3.1) and catches of lemon sole are relatively high in the IBTS. It should be noted here that for the IBTS the gear in use was fully standardized since 1983. Therefore, index data before this year (although available) should be interpreted with caution. Figure 23.3.1 displays the distribution of lemon sole in the greater North Sea obtained from IBTS Q1 data in 2016, in which year the stock was widely distributed across the central North Sea.

### 23.4 Analysis of stock trends / assessment

In 2013, lemon sole was assessed within the Working Group on the Assessment of MoU New Species (ICES, 2013). Only landings data and survey trends were available for this species. Therefore, lemon sole was defined as a category 3 species following the ICES guidelines for data limited stocks (ICES, 2013). Consequently, the basis of the advice was a trend based assessment applying method 3.2 of the guidelines for data limited stocks:

$$
C_{y+1}=\bar{C}_{y-1, y}\left(\frac{\sum_{i=y-x}^{y-1} I_{i} / x}{\sum_{i=y-z}^{x-1} I_{i} /(z-x)}\right)
$$

Here $C_{y+1}$ is the advised catch for the next year, $\bar{C}_{y-1, y}$ is the average catch of the last three years, and $I$ is the stock index (see below). By default $x=2$ and $z=5$, and this setting was used by WGNSSK in 2016.

The index of mature biomass (Figure 23.4.1.) was calculated annually from the IBTS Q1 data. For mature biomass index, the total weight per hour by centimetre length group was calculated using the length-weight relationship from Bedford et al. (1986):
$W=0.00756 L^{3.142}$.
The length-maturity ogive (Figure 23.4.2.) was then applied to calculate the mature biomass index. Lemon sole are reported to spawn in the west central North Sea during the period May to November with peak spawning during July-August (Rae, 1965). Therefore most spawning occurs between the Q1 and Q3 IBTS surveys. For this reason, the maturity ogive shown in Figure 23.4 .2 was derived from the age at maturity data (2006-2012) from both of these surveys (see stock annex for maturity-length key). Information from the spawning time would improve the accuracy of these estimates.

From WGNSSK 2015 the 2:3 ratio of the abundance index was estimated to be $-9 \%$, resulting in an advised decrease in catch from 4833 t (the average of 2013 and 2014) to 4399 t in 2016. The perception of the stock remains unchanged at WGNSSK 2016 with the discard rate and the estimate of SSB in 2015 being similar to that of 2013 and 2014 (Figure 23.4.1). Therefore, WGNSSK concludes that the extant biennial advice remains appropriate.

### 23.5 Conclusions

Discard estimates were made available for the lemon sole stock for the first time in 2015. These make the application of the DLS 3.2 advice method more justifiable than was the case when only landings data were provided. However, the new data are only available from 2013 onwards, and further backwards extension of the discard timeseries would be beneficial.

The use of only the IBTS Q1 survey is also a limitation, as the gear used is not optimum for catching flatfish and there may thus be catchability problems. Information on stock structure, biological data and catch at age information would be needed to be able to perform an analytic assessment. Age readings and maturity status evaluation techniques are still uncertain and under development.

WGNSSK concludes, on the basis of available catch and survey data, that the extent biennial advice (for 2016 and 2017) remains appropriate.

### 23.6 References

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Rae, B.B. (1965). The lemon sole. Fishing News (Books) Ltd. London, 1965. 106pp.

Table 23.6.1 Official lemon sole landings by area.

| Year | 3.4 | 4 | 7.D | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1950 | 307 | 3754 | 208 | 4269 |
| 1951 | 248 | 4710 | 314 | 5272 |
| 1952 | 243 | 4922 | 298 | 5463 |
| 1953 | 132 | 5440 | 386 | 5958 |
| 1954 | 128 | 3972 | 534 | 4634 |
| 1955 | 102 | 3836 | 141 | 4079 |
| 1956 | 96 | 3395 | 103 | 3594 |
| 1957 | 78 | 3419 | 102 | 3599 |
| 1958 | 94 | 3104 | 82 | 3280 |
| 1959 | 130 | 3647 | 82 | 3859 |
| 1960 | 153 | 4035 | 66 | 4254 |
| 1961 | 161 | 4900 | 108 | 5169 |
| 1962 | 93 | 4630 | 101 | 4824 |
| 1963 | 99 | 3791 | 66 | 3956 |
| 1964 | 134 | 4121 | 77 | 4332 |
| 1965 | 164 | 4949 | 105 | 5218 |
| 1966 | 159 | 5415 | 201 | 5775 |
| 1967 | 191 | 6188 | 331 | 6710 |
| 1968 | 185 | 6270 | 337 | 6792 |
| 1969 | 215 | 4470 | 315 | 5000 |
| 1970 | 169 | 3434 | 256 | 3859 |
| 1971 | 173 | 3967 | 357 | 4497 |
| 1972 | 168 | 3672 | 475 | 4315 |
| 1973 | 214 | 4568 | 451 | 5233 |
| 1974 | 183 | 4227 | 351 | 4761 |
| 1975 | 317 | 5029 | 33 | 5379 |
| 1976 | 361 | 4830 | 42 | 5233 |
| 1977 | 627 | 5661 | 36 | 6324 |
| 1978 | 705 | 6108 | 139 | 6952 |
| 1979 | 833 | 6428 | 260 | 7521 |
| 1980 | 722 | 6424 | 152 | 7298 |
| 1981 | 793 | 5933 | 290 | 7016 |
| 1982 | 735 | 7168 | 584 | 8487 |
| 1983 | 759 | 8257 | 491 | 9507 |
| 1984 | 595 | 6930 | 586 | 8111 |
| 1985 | 793 | 6435 | 347 | 7575 |
| 1986 | 639 | 5047 | 251 | 5937 |
| 1987 | 669 | 5516 | 310 | 6495 |
| 1988 | 642 | 5898 | 258 | 6798 |
| 1989 | 693 | 5967 | 364 | 7024 |
| 1990 | 872 | 6190 | 423 | 7485 |
| 1991 | 734 | 6618 | 428 | 7780 |
| 1992 | 952 | 6126 | 364 | 7442 |


| Year | 3.A | 4 | 7.D | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1993 | 1152 | 5839 | 422 | 7413 |
| 1994 | 801 | 5262 | 695 | 6758 |
| 1995 | 712 | 4712 | 877 | 6301 |
| 1996 | 634 | 4737 | 1151 | 6522 |
| 1997 | 766 | 4727 | 563 | 6056 |
| 1998 | 865 | 6466 | 346 | 7677 |
| 1999 | 841 | 6316 | 140 | 7297 |
| 2000 | 802 | 5980 | 388 | 7170 |
| 2001 | 583 | 5389 | 483 | 6455 |
| 2002 | 518 | 3827 | 474 | 4819 |
| 2003 | 537 | 3688 | 491 | 4716 |
| 2004 | 602 | 3543 | 424 | 4569 |
| 2005 | 669 | 3444 | 350 | 4463 |
| 2006 | 417 | 3627 | 246 | 4290 |
| 2007 | 432 | 3892 | 164 | 4488 |
| 2008 | 276 | 3465 | 234 | 3975 |
| 2009 | 262 | 2691 | 441 | 3394 |
| 2010 | 351 | 2627 | 223 | 3201 |
| 2011 | 254 | 3365 | 403 | 4022 |
| 2012 | 483 | 3084 | 459 | 4026 |
| 2013 | 290 | 2980 | 491 | 3761 |
| 2014 | 315 | 3017 | 357 | 3689 |
| 2015 | 269 | 2873 | 253 | 3394 |

Table 23.6.2 Official lemon sole landings in area 7.d by country.

| Year | BEL | DNK | FRA | NED | UK | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 10 | 0 | 174 | 0 | 24 | 0 | 208 |
| 1951 | 5 | 0 | 262 | 0 | 47 | 0 | 314 |
| 1952 | 10 | 0 | 188 | 0 | 100 | 0 | 298 |
| 1953 | 7 | 0 | 196 | 0 | 183 | 0 | 386 |
| 1954 | 9 | 0 | 361 | 0 | 164 | 0 | 534 |
| 1955 | 9 | 0 | 0 | 0 | 132 | 0 | 141 |
| 1956 | 4 | 0 | 0 | 0 | 99 | 0 | 103 |
| 1957 | 7 | 0 | 0 | 0 | 95 | 0 | 102 |
| 1958 | 1 | 0 | 0 | 0 | 81 | 0 | 82 |
| 1959 | 2 | 0 | 0 | 0 | 80 | 0 | 82 |
| 1960 | 4 | 0 | 0 | 0 | 62 | 0 | 66 |
| 1961 | 1 | 0 | 0 | 0 | 106 | 1 | 108 |
| 1962 | 2 | 0 | 0 | 0 | 99 | 0 | 101 |
| 1963 | 3 | 0 | 0 | 0 | 63 | 0 | 66 |
| 1964 | 5 | 0 | 0 | 0 | 72 | 0 | 77 |
| 1965 | 16 | 0 | 0 | 0 | 89 | 0 | 105 |
| 1966 | 7 | 0 | 0 | 0 | 194 | 0 | 201 |
| 1967 | 6 | 0 | 0 | 0 | 325 | 0 | 331 |
| 1968 | 8 | 0 | 0 | 0 | 329 | 0 | 337 |
| 1969 | 12 | 0 | 0 | 0 | 303 | 0 | 315 |
| 1970 | 16 | 0 | 0 | 0 | 240 | 0 | 256 |
| 1971 | 22 | 0 | 0 | 0 | 335 | 0 | 357 |
| 1972 | 18 | 0 | 0 | 0 | 457 | 0 | 475 |
| 1973 | 25 | 0 | 0 | 0 | 426 | 0 | 451 |
| 1974 | 16 | 0 | 0 | 1 | 334 | 0 | 351 |
| 1975 | 19 | 0 | 0 | 0 | 14 | 0 | 33 |
| 1976 | 24 | 0 | 0 | 0 | 18 | 0 | 42 |
| 1977 | 21 | 1 | 0 | 0 | 15 | 0 | 37 |
| 1978 | 45 | 2 | 63 | 0 | 31 | 0 | 141 |
| 1979 | 60 | 0 | 165 | 0 | 35 | 0 | 260 |
| 1980 | 33 | 0 | 109 | 0 | 10 | 0 | 152 |
| 1981 | 66 | 0 | 212 | 0 | 12 | 0 | 290 |
| 1982 | 96 | 0 | 406 | 1 | 81 | 0 | 584 |
| 1983 | 108 | 0 | 298 | 0 | 85 | 0 | 491 |
| 1984 | 110 | 0 | 367 | 0 | 109 | 0 | 586 |
| 1985 | 117 | 0 | 164 | 0 | 66 | 0 | 347 |
| 1986 | 77 | 0 | 133 | 0 | 41 | 0 | 251 |
| 1987 | 81 | 0 | 185 | 0 | 44 | 0 | 310 |
| 1988 | 74 | 0 | 155 | 0 | 29 | 0 | 258 |
| 1989 | 68 | 0 | 252 | 0 | 44 | 0 | 364 |
| 1990 | 68 | 0 | 272 | 0 | 83 | 0 | 423 |
| 1991 | 83 | 0 | 272 | 0 | 73 | 0 | 428 |
| 1992 | 66 | 0 | 176 | 0 | 122 | 0 | 364 |


| Year | BEL | DNK | FRA | NED | UK | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 36 | 0 | 311 | 0 | 75 | 0 | 422 |
| 1994 | 97 | 0 | 505 | 0 | 93 | 0 | 695 |
| 1995 | 138 | 0 | 584 | 0 | 155 | 0 | 877 |
| 1996 | 213 | 0 | 720 | 0 | 218 | 0 | 1151 |
| 1997 | 143 | 0 | 305 | 0 | 115 | 0 | 563 |
| 1998 | 53 | 0 | 198 | 0 | 95 | 0 | 346 |
| 1999 | 50 | 0 | 0 | 0 | 90 | 0 | 140 |
| 2000 | 62 | 0 | 200 | 0 | 126 | 0 | 388 |
| 2001 | 104 | 0 | 191 | 0 | 188 | 0 | 483 |
| 2002 | 101 | 0 | 256 | 0 | 117 | 0 | 474 |
| 2003 | 128 | 0 | 251 | 0 | 112 | 0 | 491 |
| 2004 | 120 | 0 | 198 | 1 | 105 | 0 | 424 |
| 2005 | 90 | 0 | 187 | 2 | 71 | 0 | 350 |
| 2006 | 98 | 0 | 100 | 0 | 48 | 0 | 246 |
| 2007 | 70 | 0 | 72 | 1 | 21 | 0 | 164 |
| 2008 | 140 | 0 | 46 | 3 | 45 | 0 | 234 |
| 2009 | 149 | 0 | 176 | 9 | 108 | 0 | 442 |
| 2010 | 101 | 0 | 85 | 5 | 32 | 0 | 223 |
| 2011 | 153 | 0 | 178 | 15 | 57 | 0 | 403 |
| 2012 | 171 | 0 | 167 | 20 | 0 | 0 | 358 |
| 2013 | 176 | 0 | 179 | 26 | 110 | 0 | 491 |
| 2014 | 162 | 0 | 108 | 14 | 72 | 0 | 357 |
| 2015 | 123 | 0 | 84 | 5 | 41 | 0 | 253 |

Table. 23.6.3 Official lemon sole landings in ICES sub-area 4 by country.

| Year | BEL | DNK | FRA | GER | NED | NOR | UK | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 112 | 435 | 139 | 31 | 156 | 0 | 2855 | 26 | 3754 |
| 1951 | 115 | 845 | 90 | 21 | 167 | 0 | 3430 | 42 | 4710 |
| 1952 | 98 | 391 | 227 | 26 | 168 | 0 | 3953 | 59 | 4922 |
| 1953 | 73 | 409 | 189 | 18 | 132 | 0 | 4590 | 29 | 5440 |
| 1954 | 2 | 272 | 177 | 24 | 112 | 0 | 3368 | 17 | 3972 |
| 1955 | 49 | 311 | 0 | 15 | 78 | 0 | 3374 | 9 | 3836 |
| 1956 | 48 | 222 | 0 | 19 | 58 | 0 | 3034 | 14 | 3395 |
| 1957 | 39 | 249 | 0 | 24 | 64 | 0 | 3032 | 11 | 3419 |
| 1958 | 30 | 171 | 0 | 13 | 43 | 0 | 2835 | 12 | 3104 |
| 1959 | 85 | 242 | 0 | 40 | 43 | 0 | 3226 | 11 | 3647 |
| 1960 | 155 | 577 | 0 | 46 | 67 | 0 | 3178 | 12 | 4035 |
| 1961 | 286 | 488 | 0 | 79 | 102 | 0 | 3934 | 11 | 4900 |
| 1962 | 175 | 501 | 0 | 54 | 106 | 0 | 3794 | 0 | 4630 |
| 1963 | 365 | 222 | 0 | 36 | 71 | 0 | 3097 | 0 | 3791 |
| 1964 | 484 | 358 | 0 | 62 | 75 | 0 | 3142 | 0 | 4121 |
| 1965 | 562 | 385 | 0 | 91 | 93 | 0 | 3818 | 0 | 4949 |
| 1966 | 594 | 548 | 0 | 98 | 65 | 0 | 4110 | 0 | 5415 |
| 1967 | 601 | 791 | 0 | 136 | 61 | 0 | 4599 | 0 | 6188 |
| 1968 | 422 | 775 | 0 | 96 | 34 | 0 | 4943 | 0 | 6270 |
| 1969 | 292 | 639 | 0 | 80 | 36 | 0 | 3423 | 0 | 4470 |
| 1970 | 241 | 307 | 0 | 52 | 58 | 0 | 2776 | 0 | 3434 |
| 1971 | 348 | 514 | 0 | 54 | 122 | 0 | 2929 | 0 | 3967 |
| 1972 | 423 | 530 | 0 | 59 | 130 | 0 | 2530 | 0 | 3672 |
| 1973 | 566 | 478 | 0 | 73 | 217 | 16 | 3218 | 0 | 4568 |
| 1974 | 486 | 447 | 0 | 59 | 269 | 0 | 2966 | 0 | 4227 |
| 1975 | 748 | 521 | 0 | 83 | 299 | 0 | 3367 | 11 | 5029 |
| 1976 | 493 | 506 | 0 | 68 | 308 | 0 | 3443 | 12 | 4830 |
| 1977 | 618 | 321 | 0 | 71 | 262 | 0 | 4387 | 2 | 5661 |
| 1978 | 760 | 517 | 28 | 54 | 231 | 0 | 4518 | 0 | 6108 |
| 1979 | 674 | 876 | 136 | 41 | 390 | 0 | 4308 | 3 | 6428 |
| 1980 | 484 | 599 | 102 | 49 | 303 | 0 | 4885 | 2 | 6424 |
| 1981 | 555 | 605 | 237 | 39 | 412 | 0 | 4084 | 1 | 5933 |
| 1982 | 879 | 670 | 419 | 52 | 759 | 0 | 4386 | 3 | 7168 |
| 1983 | 1122 | 735 | 402 | 28 | 1009 | 0 | 4957 | 4 | 8257 |
| 1984 | 1144 | 567 | 344 | 22 | 0 | 0 | 4850 | 3 | 6930 |
| 1985 | 989 | 555 | 157 | 26 | 0 | 0 | 4703 | 5 | 6435 |
| 1986 | 511 | 577 | 103 | 16 | 0 | 0 | 3839 | 1 | 5047 |
| 1987 | 448 | 742 | 174 | 14 | 0 | 0 | 4137 | 1 | 5516 |
| 1988 | 539 | 639 | 184 | 14 | 301 | 0 | 4220 | 1 | 5898 |
| 1989 | 441 | 828 | 176 | 40 | 397 | 0 | 4083 | 2 | 5967 |
| 1990 | 491 | 1007 | 208 | 49 | 0 | 0 | 4431 | 4 | 6190 |
| 1991 | 544 | 1099 | 250 | 41 | 0 | 12 | 4666 | 6 | 6618 |
| 1992 | 577 | 1149 | 177 | 30 | 0 | 13 | 4175 | 5 | 6126 |


| Year | BEL | DNK | FRA | GER | NED | NOR | UK | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 525 | 966 | 240 | 37 | 0 | 9 | 4059 | 3 | 5839 |
| 1994 | 436 | 597 | 436 | 27 | 0 | 11 | 3754 | 1 | 5262 |
| 1995 | 588 | 585 | 412 | 70 | 0 | 9 | 3046 | 2 | 4712 |
| 1996 | 592 | 547 | 534 | 67 | 0 | 18 | 2976 | 3 | 4737 |
| 1997 | 504 | 499 | 224 | 76 | 0 | 29 | 3391 | 4 | 4727 |
| 1998 | 815 | 796 | 197 | 149 | 838 | 23 | 3643 | 5 | 6466 |
| 1999 | 662 | 1015 | 0 | 62 | 681 | 24 | 3866 | 6 | 6316 |
| 2000 | 711 | 1277 | 184 | 72 | 492 | 17 | 3222 | 5 | 5980 |
| 2001 | 694 | 1281 | 191 | 77 | 451 | 22 | 2666 | 7 | 5389 |
| 2002 | 604 | 971 | 190 | 116 | 402 | 17 | 1521 | 6 | 3827 |
| 2003 | 517 | 1008 | 239 | 136 | 369 | 16 | 1399 | 4 | 3688 |
| 2004 | 667 | 1113 | 120 | 81 | 355 | 12 | 1192 | 3 | 3543 |
| 2005 | 595 | 1057 | 102 | 85 | 402 | 13 | 1188 | 2 | 3444 |
| 2006 | 552 | 968 | 57 | 183 | 412 | 13 | 1440 | 2 | 3627 |
| 2007 | 542 | 1136 | 65 | 143 | 367 | 23 | 1610 | 6 | 3892 |
| 2008 | 527 | 925 | 47 | 120 | 434 | 26 | 1383 | 4 | 3466 |
| 2009 | 389 | 898 | 88 | 64 | 294 | 31 | 927 | 2 | 2693 |
| 2010 | 375 | 821 | 32 | 102 | 323 | 35 | 935 | 2 | 2625 |
| 2011 | 387 | 999 | 56 | 96 | 641 | 27 | 1157 | 2 | 3365 |
| 2012 | 406 | 999 | 34 | 61 | 587 | 30 | 0 | 2 | 2119 |
| 2013 | 527 | 649 | 27 | 67 | 479 | 16 | 1214 | 2 | 2981 |
| 2014 | 648 | 626 | 27 | 63 | 425 | 23 | 1202 | 3 | 3017 |
| 2015 | 425 | 794 | 16 | 82 | 423 | 12 | 1116 | 3 | 2873 |

Table 23.6.4 Official landings in area 3.a by country.

| Year | BEL | DNK | GER | NED | SWE | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0 | 100 | 1 | 0 | 206 | 0 | 307 |
| 1951 | 0 | 74 | 1 | 0 | 173 | 0 | 248 |
| 1952 | 0 | 64 | 0 | 0 | 179 | 0 | 243 |
| 1953 | 0 | 35 | 0 | 0 | 97 | 0 | 132 |
| 1954 | 0 | 33 | 0 | 0 | 95 | 0 | 128 |
| 1955 | 0 | 29 | 0 | 0 | 73 | 0 | 102 |
| 1956 | 0 | 33 | 0 | 0 | 63 | 0 | 96 |
| 1957 | 0 | 27 | 0 | 0 | 51 | 0 | 78 |
| 1958 | 0 | 38 | 0 | 0 | 56 | 0 | 94 |
| 1959 | 0 | 71 | 0 | 0 | 59 | 0 | 130 |
| 1960 | 0 | 95 | 1 | 0 | 57 | 0 | 153 |
| 1961 | 0 | 90 | 0 | 0 | 71 | 0 | 161 |
| 1962 | 0 | 92 | 1 | 0 | 0 | 0 | 93 |
| 1963 | 0 | 99 | 0 | 0 | 0 | 0 | 99 |
| 1964 | 0 | 133 | 1 | 0 | 0 | 0 | 134 |
| 1965 | 0 | 163 | 1 | 0 | 0 | 0 | 164 |
| 1966 | 0 | 159 | 0 | 0 | 0 | 0 | 159 |
| 1967 | 0 | 189 | 1 | 0 | 0 | 1 | 191 |
| 1968 | 0 | 184 | 0 | 0 | 0 | 1 | 185 |
| 1969 | 0 | 215 | 0 | 0 | 0 | 0 | 215 |
| 1970 | 0 | 169 | 0 | 0 | 0 | 0 | 169 |
| 1971 | 0 | 173 | 0 | 0 | 0 | 0 | 173 |
| 1972 | 0 | 168 | 0 | 0 | 0 | 0 | 168 |
| 1973 | 0 | 214 | 0 | 0 | 0 | 0 | 214 |
| 1974 | 0 | 183 | 0 | 0 | 0 | 0 | 183 |
| 1975 | 0 | 263 | 1 | 1 | 52 | 0 | 317 |
| 1976 | 10 | 294 | 1 | 19 | 37 | 0 | 361 |
| 1977 | 9 | 528 | 2 | 37 | 51 | 0 | 627 |
| 1978 | 4 | 628 | 2 | 12 | 59 | 0 | 705 |
| 1979 | 7 | 704 | 1 | 10 | 111 | 0 | 833 |
| 1980 | 12 | 622 | 0 | 0 | 87 | 1 | 722 |
| 1981 | 1 | 710 | 0 | 3 | 75 | 4 | 793 |
| 1982 | 2 | 647 | 0 | 9 | 77 | 0 | 735 |
| 1983 | 3 | 636 | 0 | 10 | 110 | 0 | 759 |
| 1984 | 6 | 525 | 0 | 0 | 64 | 0 | 595 |
| 1985 | 0 | 729 | 0 | 0 | 64 | 0 | 793 |
| 1986 | 7 | 576 | 0 | 0 | 56 | 0 | 639 |
| 1987 | 24 | 577 | 0 | 0 | 68 | 0 | 669 |
| 1988 | 11 | 569 | 0 | 6 | 56 | 0 | 642 |
| 1989 | 8 | 610 | 0 | 0 | 75 | 0 | 693 |
| 1990 | 16 | 782 | 0 | 0 | 74 | 0 | 872 |
| 1991 | 11 | 640 | 0 | 0 | 83 | 0 | 734 |
| 1992 | 22 | 793 | 0 | 0 | 120 | 17 | 952 |


| Year | BEL | DNK | GER | NED | SWE | OTHER | TOTAL |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1993 | 14 | 980 | 4 | 0 | 141 | 17 | 1156 |
| 1994 | 10 | 648 | 2 | 0 | 127 | 16 | 803 |
| 1995 | 27 | 576 | 2 | 0 | 91 | 18 | 714 |
| 1996 | 0 | 513 | 1 | 0 | 97 | 24 | 635 |
| 1997 | 0 | 628 | 2 | 0 | 115 | 23 | 768 |
| 1998 | 0 | 743 | 3 | 0 | 100 | 22 | 868 |
| 1999 | 0 | 731 | 3 | 0 | 88 | 22 | 844 |
| 2000 | 0 | 722 | 1 | 0 | 65 | 15 | 803 |
| 2001 | 0 | 511 | 1 | 0 | 53 | 19 | 584 |
| 2002 | 0 | 457 | 4 | 0 | 41 | 20 | 522 |
| 2003 | 0 | 451 | 6 | 30 | 35 | 21 | 543 |
| 2004 | 0 | 472 | 5 | 82 | 29 | 19 | 607 |
| 2005 | 0 | 468 | 5 | 147 | 38 | 16 | 674 |
| 2006 | 0 | 321 | 8 | 40 | 32 | 16 | 417 |
| 2007 | 0 | 374 | 5 | 16 | 18 | 19 | 432 |
| 2008 | 0 | 239 | 7 | 3 | 15 | 12 | 276 |
| 2009 | 0 | 233 | 4 | 1 | 15 | 9 | 262 |
| 2010 | 0 | 286 | 3 | 35 | 19 | 7 | 350 |
| 2011 | 0 | 223 | 0 | 0 | 12 | 16 | 254 |
| 2012 | 0 | 446 | 3 | 0 | 15 | 18 | 482 |
| 2013 | 0 | 259 | 3 | 5 | 10 | 12 | 289 |
| 2014 | 0 | 276 | 7 | 12 | 14 | 6 | 315 |
| 2015 | 0 | 250 | 4 | 0 | 9 | 6 | 269 |
|  |  |  |  |  |  |  |  |

Table 23.6.5 Mature biomass index (g/hour) calculated from IBTS Q1 data by WGNSSK in 2016.

|  | Year | Biomass index |
| :--- | :--- | :--- |
| 1983 | 1.61 | 1.629 |
| 1984 | 1.273 |  |
| 1985 | 1.467 |  |
| 1986 | 1.313 |  |
| 1987 | 1.357 |  |
| 1988 | 1.583 |  |
| 1989 | 1.548 |  |
| 1990 | 1.171 |  |
| 1991 | 1.542 |  |
| 1992 | 1.927 |  |
| 1993 | 1.185 |  |
| 1994 | 1.157 |  |
| 1995 | 1.381 |  |
| 1996 | 1.223 |  |
| 1997 | 1.733 |  |
| 1998 | 1.787 |  |
| 1999 | 1.702 |  |
| 2000 | 1.377 |  |
| 2001 | 1.819 |  |
| 2002 | 1.707 |  |
| 2003 | 1.683 |  |
| 2004 | 1.22 |  |
| 2005 | 1.02 |  |
| 2006 | 1.331 |  |
| 2007 | 1.331 |  |
| 2008 | 0.862 |  |
| 20109 | 0.954 |  |
| 2010 | 1.265 |  |
|  | 1.895 |  |
|  | 1.249 |  |
|  | 0.968 |  |



Figure 23.2.1. Lemon sole in Subarea 4, and Divisions 3.a and 7.d. Official landings (tonnes) of lemon sole by area in the greater North Sea, 1950-2015.


Figure 23.2.2. Lemon sole in Subarea 4, and Divisions 3.a and 7.d. Official landings (tonnes) of lemon sole in area 7.d by country for 1950-2015. Note that official landings data for UK are missing for 2012.


Figure 23.2.3. Lemon sole in Subarea 4, and Divisions 3.a and 7.d. Official landings (tonnes) of lemon sole in area 4 by country, for 1950-2015. Note that official landings data for UK are missing for 2012.


Figure 23.2.4. Lemon sole in Subarea 4, and Divisions 3.a and 7.d. Official landings (tonnes) of lemon sole in area 3.a by country, for 1950-2015.


Figure 23.2.5. Lemon sole in Subarea 4, and Divisions 3.a and 7.d. Time-series of official landings (red dots) along with WG estimates of landings (dark bars) and discards (light bars). Note that the discard-rate estimate for 2012 was based on data submissions from a small subset of countries only, and is unlikely to be representative.
lemon sole 2015: all lengths


Figure 23.3.1. Distribution of lemon sole in the greater North Sea derived from IBTS Q1 data (2015).


Figure 23.4.1. Index of mature biomass ( $\mathrm{kg} / \mathrm{hr}$ )for Subarea 4 derived from IBTS Q1 data


Figure 23.4.2. Length based maturity ogive for Lemon sole derived from IBTS Q1 data.

## 23 Witch (Glyptocephalus cynoglossus) in Subarea 4 and Divisions 3.a and 7.d (North Sea, Skagerrak and Kattegat, Eastern English Channel)

### 23.1 General

Witch flounder (Glyptocephalus cynoglossus) was assessed, between 2010 and 2013, by the Working Group on Assessment of New MoU Species (WGNEW, ICES 2013.a). Since 2014 WGNEW was dissolved thus this species 'was included in the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK).
Following the ICES guidelines for data limited stocks (ICES, 2012) witch was defined as a category 3 species as only official landings and survey data were available. The biennial advice, drafted in 2013 (ICES, 2013b), was based on stock size indicators (standardized CPUE in number/hour) derived from IBTS (both Q1 and Q3) and exploratory estimates (merely indicative of trends and not used for catch forecast) suggesting that fishing mortality is above potential $\mathrm{F}_{\text {MSY }}$ proxies. In 2015, witch flounder was included into the official data call for the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK) and the biennial advice was evaluated by this group. The new data call for the WGNSSK 2016 included landing and discard data for the years 2012-2015 for attempting to give catch advice for this species.

### 23.1.1 Biology and ecosystem aspects

The existing knowledge of witch biology is summarized in the Stock Annex.
In 2009, witch flounder has been included as a mandatory species in the EU Data Collection Framework (DCF). Accordingly, Denmark and Sweden started the regular sampling of biological data, i.e. length, weight, maturity status and age, in 3 .a and 4 both in discards and landings. Scotland has also been collecting biological samples since 2009 but only from the landings.

Age readings techniques are now well established while the macroscopic evaluation of maturity status is still uncertain and a histological analysis of the gonads is under development and it is planned to be ready before the benchmark scheduled in 2017.

### 23.1.2 Management regulations

According to EU-Regulations a precautionary TAC is given in EU waters of 3.a and 4 together with lemon sole (Microstomus kitt). The TACs have been stable, varying around 6000 t since 2006. There is no official Minimum Landing Size (MLS) specified in EU waters. However, in most of the countries reporting catches the landing of witch below 28 cm is prohibited. Currently, lemon sole and witch flounder are managed under a combined species TAC, which prevents the effective control of the single species exploitation rates and could potentially lead to the overexploitation of either species. Furthermore, witch flounder is mainly a bycatch species in a mixed fisheries (although some limited seasonal target fisheries occurs) thus a TAC alone may not be appropriate as a management tool.

### 23.2 Fisheries data

### 23.2.1 Historical landings

North Sea witch flounder's landings have declined in the last decade, but from 2011 a general increasing trend is observed. This species is nowadays mainly landed by Denmark, Norway, Sweden and Germany in both areas (3.a and 4) and UK mainly in Subarea 4. The Netherlands reports only a small fraction of the total landings in subarea 4 as this species it is mostly discarded. In division 3.a, Denmark is landing the largest amount of witch flounder, while in Subarea 4 it is Scotland having the largest portion of the landings.

### 23.2.2 InterCatch

In 2014, witch flounder was included for the first time into the data call for WGNSSK 2014 and since 2015 the data call was extended to obtain landing and discard data for the years 2012-2015. From all countries data were uploaded to the InterCatch data portal. Norway did not report any discards.

Discards could thus be raised for the period 2012-2015 and catches estimated. In general, the discard rate is moderately low and it has been decreasing from $23 \%$ in 2012 to 11 and $10 \%$ respectively in 2013 and 2014 and increased again in 2015 (18\%). However, it should be noted that not all metiers were sampled in every quarter and that raising procedure may not be adequate in all cases. Thus for some metiers the applied raising procedure might introduce some bias to the total discard estimates. An overview of the reported landings and discards and the resulting discard rates for all fleets is given in table 24.4.1. Landings showed a slight decrease from 2014 to 2015, around 2300t.

For 2015, the largest amount of landings and discards was reported by Denmark in Division 3.a using the OTB_CRU_90-119_0_0_all metier and by Scotland in Subarea 4 using the OTB_DEF_>=120_0_0_all métier (Figures 26.2.2.1-3). The total catch estimated with InterCatch was 2649 t , of which only 410 t were discards ( $18 \%$ of total catch).

### 23.3 Survey data / recruit series

The International Bottom Trawl Survey (IBTS) performed every year during the first and third quarter since 1975 provides indices for the North Sea and 3.a. Furthermore a time series of Dutch Beam Trawl Survey (BTS) data (1985-2008) in 4 is also available but it was not explored during the current assessment. The IBTS seem to be the most valuable and promising data source to be used as tuning fleet for the assessment, particularly during Q1 when more stations are usually fished and the time series is longer (Figure 24.3.1).

### 23.4 Analysis of stock trends / assessment

Witch flounder has been classified as category 3 stocks following the guidelines of the ICES Data Limited Stocks (DLS) methodological document (ICES 2012). This category includes stocks for which survey indices (or other indicators of stock size) are available and provide reliable indications of trends in stock metrics.

Consequently, the basis of the biennial advice in 2013 was a trend based assessment applying method 3.2 of the guidelines for data limited stocks:

$$
C_{y+1}=C_{y-1}\left(\frac{\sum_{i=y-x}^{y=y-1} I_{i} / x}{\sum_{i=y-z}^{y-x-1} I_{i} /(z-x)}\right)
$$

Where $C_{y+1}$ is the advised catch for the next year, $C_{y-1}$ should be the average catch of the last three years, and $I$ is the stock index. By default $x=2$ and $z=5$. A mature biomass index in kg per hour was estimated from the IBTS Q1 and Q3 survey. The choice to compare three versus five rather than two versus three years index values applied for the advice 2013 was made for accounting the inter-annual variability of surveys. Recent more detailed analysis of the gonads (i.e. ongoing work at the Swedish Institute for Marine Research) revealed that this species becomes reproductively mature at age 5 and therefore considering a three versus five years average will include at least one generation.

A logistic regression applied in 2014 on the DATRAS CA records showed that L50, i.e. the length at which $50 \%$ of the stock is mature, corresponds to 34 cm (ICES 2015). Thus, as in 2015, the mature biomass indices were estimated including all specimens larger than 34 cm and the LW relationship as in 2015 was used (ICES 2015).

For IBTS Q1 survey the three most recent year indices of the mature biomass (in kg per hour) (2014-2016) were more than $300 \%$ higher than five previous year index although it declined in 2016, while for IBTS Q3, the three most recent year indices (2013-2015) were more than $40 \%$ higher than five previous year index.

During WGNSSK 2016, a mature biomass index in kg per hour as derived from both surveys (IBTS Q1 and Q3) was estimated in accordance with the DLS guidelines and thus the mean of the two most recent year (2015-2016 and 2014-2015 for IBTS Q1 and Q3, respectively) index was compared to the mean of the three previous years (20122014 and 2011-2013 for IBTS Q1 and Q3, respectively) indices. The stock size indicator (i.e. the mature biomass index) in the last two years was $3 \%$ higher for IBTS Q1 but about $17 \%$ less for IBTS Q3 than the average of the three previous years.

Based on these information, WGNSSK 2016 consider that the biannual advice issued by ICES in 2015 and valid for 2016 and 2017 should be maintained and total catches in 2017 should be no more than 3107 tonnes.

### 23.5 References

ICES 2013.a. Report of the Working Group on Assessment of New MoU Species (WGNEW), 2428 March 2013, ICES Headquarters, Denmark. ICES CM 2013/ACOM.

ICES 2013b. Witch in Subarea IV and Division 3.a and VIId, Report of the ICES Advisory Committee, 2013. ICES Advice, 2013. Book 6, Section 6.4.35.

ICES 2012. ICES implementation of advice for data limited stocks in 2012. Report in support of ICES advice. ICES CM2012/ACOM:68.

ICES 2015. Report of the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), 28 April-7 May, ICES HQ, Copenhagen, Denmark. ICES CM 2015/ACOM:13. 1229 pp.

Table 24.4.1. Witch flounder in area 4 and division 3.a. Summary of the assessment. Landings, discards and catches are in tonnes. The IBTS indices indicate mature biomass in $\mathbf{k g} / \mathrm{hour}$.

| Year | Official LANDINGS | $\begin{gathered} \text { ICES } \\ \text { LANDINGS } \end{gathered}$ | $\begin{aligned} & \text { ICES } \\ & \text { CATCHES } \end{aligned}$ | $\begin{gathered} \text { ICES } \\ \text { DISCARDS } \end{gathered}$ | IBTS Q1 Index | IBTS Q3 index | Discard rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 1174 |  |  |  | 0.08 |  |  |
| 1969 | 891 |  |  |  | 0.04 |  |  |
| 1970 | 597 |  |  |  | 0.15 |  |  |
| 1971 | 843 |  |  |  | 0.01 |  |  |
| 1972 | 908 |  |  |  | 0.01 |  |  |
| 1973 | 1494 |  |  |  | 0.06 |  |  |
| 1974 | 1138 |  |  |  | 0.04 |  |  |
| 1975 | 1841 |  |  |  | 0.03 |  |  |
| 1976 | 1496 |  |  |  | 0.13 |  |  |
| 1977 | 1618 |  |  |  | 0.04 |  |  |
| 1978 | 1664 |  |  |  | 0.05 |  |  |
| 1979 | 1572 |  |  |  | 0.07 |  |  |
| 1980 | 1883 |  |  |  | 0.03 |  |  |
| 1981 | 1933 |  |  |  | 0.38 |  |  |
| 1982 | 3155 |  |  |  | 0.06 |  |  |
| 1983 | 3606 |  |  |  | 0.15 |  |  |
| 1984 | 3903 |  |  |  | 0.11 |  |  |
| 1985 | 3979 |  |  |  | 0.16 |  |  |
| 1986 | 3579 |  |  |  | 0.17 |  |  |
| 1987 | 3700 |  |  |  | 0.21 |  |  |
| 1988 | 3290 |  |  |  | 0.07 |  |  |
| 1989 | 3841 |  |  |  | 0.30 |  |  |
| 1990 | 3862 |  |  |  | 0.12 |  |  |
| 1991 | 3641 |  |  |  | 0.10 | 0.11 |  |
| 1992 | 3164 |  |  |  | 0.39 | 0.12 |  |
| 1993 | 2673 |  |  |  | 0.28 | 0.06 |  |
| 1994 | 2696 |  |  |  | 0.09 | 0.08 |  |
| 1995 | 2810 |  |  |  | 0.25 | 0.13 |  |
| 1996 | 2790 |  |  |  | 0.09 | 0.10 |  |
| 1997 | 3494 |  |  |  | 0.25 | 0.17 |  |
| 1998 | 3786 |  |  |  | 0.25 | 0.08 |  |
| 1999 | 4024 |  |  |  | 0.19 | 0.12 |  |
| 2000 | 4422 |  |  |  | 0.24 | 0.04 |  |
| 2001 | 4206 |  |  |  | 0.13 | 0.11 |  |
| 2002 | 3640 |  |  |  | 0.16 | 0.09 |  |
| 2003 | 3281 |  |  |  | 0.12 | 0.05 |  |
| 2004 | 3029 |  |  |  | 0.12 | 0.08 |  |
| 2005 | 2813 |  |  |  | 0.14 | 0.05 |  |
| 2006 | 2303 |  |  |  | 0.06 | 0.08 |  |
| 2007 | 2236 |  |  |  | 0.08 | 0.12 |  |
| 2008 | 1953 |  |  |  | 0.11 | 0.06 |  |
| 2009 | 1818 |  |  |  | 0.06 | 0.05 |  |
| 2010 | 1490 |  |  |  | 0.04 | 0.06 |  |
| 2011 | 1530 |  |  |  | 0.05 | 0.09 |  |
| 2012 | 1895 | 1953 | 2544 | 592 | 0.09 | 0.13 | 0.303 |
| 2013 | 1993 | 2020 | 2272 | 252 | 0.08 | 0.13 | 0.125 |
| 2014 | 2646 | 2669 | 2950 | 281 | 0.29 | 0.08 | 0.105 |
| 2015 | 2359 | 2238 | 2649 | 410 | 0.19 | 0.12 | 0.183 |
| 2016 |  |  |  |  | 0.13 |  |  |

Table 24.6.1. Witch flounder in area 4 and division 3.a. Official ICES landings by area 4 and division 3.a.

|  |  | YeAR | 3.A |
| :--- | :--- | :--- | :--- |


| Year | 3.A |  | 4 | Тот |
| :---: | :---: | :---: | :---: | :---: |
| 1992 | 1237 | 1927 |  | 3164 |
| 1993 | 950 | 1723 |  | 2673 |
| 1994 | 771 | 1925 |  | 2696 |
| 1995 | 939 | 1871 |  | 2810 |
| 1996 | 902 | 1888 |  | 2790 |
| 1997 | 1502 | 1992 |  | 3494 |
| 1998 | 1986 | 1800 |  | 3786 |
| 1999 | 2239 | 1785 |  | 4024 |
| 2000 | 2477 | 1945 |  | 4422 |
| 2001 | 1939 | 2267 |  | 4206 |
| 2002 | 2006 | 1634 |  | 3640 |
| 2003 | 1646 | 1635 |  | 3281 |
| 2004 | 1788 | 1241 |  | 3029 |
| 2005 | 1605 | 1208 |  | 2813 |
| 2006 | 1043 | 1260 |  | 2303 |
| 2007 | 949 | 1287 |  | 2236 |
| 2008 | 783 | 1170 |  | 1953 |
| 2009 | 773 | 1045 |  | 1818 |
| 2010 | 675 | 815 |  | 1490 |
| 2011 | 693 | 837 |  | 1530 |
| 2012 | 1107 | 788 |  | 1895 |
| 2013 | 1000 | 993 |  | 1993 |
| 2014 | 1562 | 1085 |  | 2646 |
| 2015 | 1282 | 956 |  | 2238 |

Landings (tonnes)


Figure 24.2.1.1. Witch flounder in area 4 and division 3.a. Total official landings (in tonnes).


Figure 24.2.2.1. Witch flounder in Subarea 4 and division 3.a. Landings by metier and country in 2015.


Figure 24.2.2.2. Witch flounder in Subarea 4 and division 3.a. Discards by metier and country in 2015. Reported discards panel (a), raised discards panel (b).


Figure 24.2.2.3. Witch flounder in Subarea 4 and division 3.a. Estimated landings and discards by countries in 2015.


Figure 24.3.1. Witch flounder in Subarea 4 and division 3.a. Aggregated distribution over the entire time series in the North Sea derived from IBTS Q1 (upper) and Q3 (lower) using data collected between 1968 and 2015.

## Annex 01 List of participants

|  | PARTICIPANTS | INSTITUTE |
| :--- | :--- | :--- |
| Jan Jaap Poos | IMARES | Netherlands |
| Ruben Verkempynck | IMARES | Netherlands |
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| Tanja Miehte | MARLAB | Scotland |
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| Clara Ulrich | DTU | Denmark |
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| Mats Ulmestrand | SLU | Sweden |
| Andreas Sundelöf | SLU | Sweden |
| Alexander Kempf | TI | Germany |
| Scott Large | SLU | DK |
| Massimiliano Cardinale (by correspondence) | Sweden |  |
|  |  |  |

## Annex 02 Recommendations

The following table summarises the main recommendations arising from the WGNSSK and identifies suggested responsibilities for action.

| Recommendation | For follow up by: |
| :--- | :--- |

WGNSSK evaluated the guidelines for reference points ACOM and made the following observations:

The current guidelines propose to use the sigma on $\ln \left(\mathrm{F}_{\mathrm{bar}}\right)$ and $\ln (\mathrm{SSB})$ from the final year in the assessment to derive $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{B}_{\mathrm{pa}}$ from $\mathrm{F}_{\mathrm{lim}}$ and $\mathrm{Blim}_{\mathrm{lim}}$. For many stocks in WGNSSK this would bring $\mathrm{B}_{\mathrm{pa}}$ and $\mathrm{F}_{\mathrm{pa}}$ closer to the limit reference points compared to the default used so far (1.4* $\mathrm{B}_{\mathrm{lim}}$ and $\mathrm{F}_{\mathrm{lim}} / 1.4$ ). Furthermore, the current guidelines do not take into account retrospective uncertainty/bias in the calculation of the uncertainty in the last assessment year, and the interpretation of the latter can vary between assessment methods because they handle it in different ways. More careful thinking is needed about what uncertainty in the final year in the assessment means. The guidelines are also still only draft and have not been finally agreed by ACOM. Until final guidelines are agreed by ACOM, WGNSSK recommends the use of the estimated sigmas from assessments only if the precautionary reference points that result from their application are larger for biomass or lower for $F$ than the ones derived by the default settings $\left(1.4^{*} \mathrm{~B}_{\mathrm{lim}}\right.$ and $F_{\text {lim }} / 1.4$ ). Such an approach is also used in the US when taking scientific uncertainty into account in the setting of buffers between the OFL and ABC (Dichmont et al. in press, Ralston et al. 2011).

A decision is needed about whether $\mathrm{F}_{\mathrm{pa}}$ and/or Fp. 05 restrict the upper limit of FMSY ranges as both reference points differ from one another in most cases. A clear decision is needed about whether the MSY approach or the precautionary approach is the leading principle behind reference points.

In general, the increasing number of reference points is worrying, also making the communication to stakeholders difficult. WGNSSK suggests to keep the number of reference points to a minimum, and to keep methodology as consistent and concise as possible (e.g. $\mathrm{F}_{\mathrm{pa}}=\mathrm{Fp} .05$ ).

The last step in the $B_{\text {trigger }}$ decision tree in the current draft guidelines is unclear. It can happen that the 5th percentile of the current SSB is below $B_{p a}$ (e.g., if there is a depleted stock and/or where there is large uncertainty in the assessment). WGNSSK suggests to make clear in the guidelines that $B_{p a}$ is the lower limit for $B_{\text {trigger }}$ in all cases.

## References:

Dichmont, C.M., Punt, A.E., Dowling, N., De Oliveira, J.A.A., Little, L.R., Sporcic, M., Fulton, E., Gorton, R., Klaer, N., Haddon, M. and D.S. Smith. In press. Is risk consistent across tier-based harvest control rule management systems? A comparison of four case-studies. Fish and Fisheries, doi: 10.1111/faf.12142.

Ralston, S., Punt, A.E., Hamel, O.S., DeVore, J. and Conser, R.J. (2011) An approach to quantifying scientific uncertainty in stock assessment. Fishery Bulletin 109, 217-231.

Underwater TV surveys for Nephrops are carried out in summer after WGNSSK. This questions an assessment in spring and leads to a high probability that the advice needs to be reopened once the latest TV survey estimates are available. WGNSSK recommends postponing the final Nephrops assessment and advice to autumn. Analyses of catch data from the previous year and a preliminary assessment can still be conducted during the spring meeting to support mixed fishery advice. But the final assessment and advice could be produced either by correspondence or in a meeting together with the Celtic Sea Nephrops stocks. However, the timing of TV surveys needs to ensure that at least in autumn the most recent data are available.

There is currently little guidance about how catch categories should be reported, and how they should be used in raising discards where discard information is not provided. This is an issue that affects all Expert Groups that have to provide catch advice, so a common approach is needed. BMS landings, observer discards and log-book recorded discards should sum up to discard data provided so far (i.e. double-counting should be avoided), and when performing raising procedures, these three categories should be combined to provide raising factors, and the raising procedure in Intercatch should be adapted as necessary. This pro-

ACOM, Scotland, England, Denmark

ACOM, ICES Data Center
vides a robust approach, independent of how countries categorize catches when providing catch data. WGNSSK recommends that ICES provides a harmonized approach across all Expert Groups.

The French FRGFS survey does not currently receive ACOM, France funding from DCF and may be discontinued. This survey provides tuning series used in the assessment of red mullet and 7d plaice. This survey needs to be maintained in order to maintain the integrity of current assessments for these stocks.

The current guidelines for reopening advice for DLS
ACOM stock is vague, referring to a "change in perception" of stock status. It is not entirely clear what this means, and what constitutes a perception that would be different enough to lead to re-opening. It is recommended that clearer guidelines are provided in this regard.

## Annex 03 ToRs for next meeting

WGNSSK - Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak

2016/2/ACOMXX The Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), chaired by José De Oliveira, UK, in meet in Copenhagen, 25 April - 4 May 2017 and by correspondence in September 2017 to:
a) Address generic ToRs for Regional and Species Working Groups. The Norway pout assessments shall be developed by correspondence;

The assessments will be carried out on the basis of the stock annex in National Laboratories, prior to the meeting. The assessments must be available for audit on the first day of the meeting.

Material and data relevant for the meeting must be available to the group no later than 11 April 2017 according to the Data Call 2017.

WGNSSK will report by 11 May 2017, and by 28 September 2017 (Norway pout) for the attention of ACOM.

## Annex 04 Update Forecasts and Assessments

### 4.1 Summary

The Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak [WGNSSK] (Chairs: Alexander Kempf, DE and Jose de Oliveira, UK) met by correspondence at the beginning of October 2016 to evaluate new information from the fisheries independent surveys carried out during 2016 subsequent to the meeting of the group in April/May. Also this year a deviation from the manual occurred during the IBTS q3 survey. It was tested whether the CPUE from 15 min . hauls does not significantly differ from 30 min . hauls. In most ICES rectangles one 30 min . and one 15 min. haul was conducted instead of two 30 min . hauls. Analyses of theIBTS working group do not show significant differences between 15 min and 30 min hauls. Therefore, all hauls were used for the IBTS q3 index calculations (relevant for cod, whiting, haddock and saithe). However, Verena Trenkel (IFREMER) has also analysed 15 and 30 min hauls in comparison and found systematic differences in CPUE values (higher in the 15 min hauls when standardized to 1 hour). This needs further investigation to derive final conclusions taking also the second year of data into account.

The WGNSSK followed the protocol defined by the Ad hoc Group on Criteria for Reopening Fisheries Advice (AGCREFA; ICES CM 2008/ACOM: 60) in its evaluation of the survey information - fitting the RCT3 regression model to data that included the 2016 survey information to estimate the recent recruitment abundance and then comparing the prediction and its associated uncertainty with the assumptions made in forecasts used as the basis for the ACOM spring advice.

As every year, the indices used in the current update must be considered as provisional and may be revised for the assessment in May next year.

An update is also presented for the Nephrops stocks, given that UTV surveys usually take place over summer. This allows for a considerably smaller time lag between the last abundance observations and their use for next year's advice.

The comparisons indicated that there is potential for re-opening of the June advice for cod in IV, IIIa and VIId, saithe in IV, IIIa and VI as well as Nephrops in FUs 6, 7 and FU 8. For Nephrops FU6, the new catch advice would be 1125 t compared to 740 t in June. If the discarding exemptions for 2016 continue to apply for 2017 then the wanted landings would be 1020 t (compared to 671 t in June). For Nephrops FU7, the new catch advice would be 11852 t compared to 6844 t in June. If the discarding exemptions for 2016 continue to apply for 2017 then the wanted landings would be 11813t (compared to 6821 t in June). The strong increase may be related to a strong recruitment. These recruits will not show up immediately in the fishery given the current selectivity. Given that the TAC has not been taken in recent years caution is needed when increasing the TAC for FU7 further. It has to be ensured this does not lead to overexploitation in other FUs. For FU 8 the catch advice would change from 2123 t to 2548 t. If the discarding exemptions for 2016 continue to apply for 2017 then the wanted landings would be 2190 t (compared to 1825 t in June). For cod in IV, IIIa and VIId the advice for total catch would change from 47431 tonnes to 47359 tonnes . Despite a considerably lower 2016 recruitment estimate, the change in TAC is small because the 2016 indices (for the older ages) lead to higher abundances for older age groups at the start of the intermediate year in the SAM modelling framework. For saithe in IV, IIIa and VI the advice would change from 116605 to 134962 tonnes. A very strong incoming year class and significantly higher age 4 index has been detected in the IBTS q3 survey. However, the age 3
survey index is highly uncertain and the internal consistency between age 3 and age 4 is extremely poor. Therefore, the perception of the incoming year class strength may change considerably in the next years. In addition age 3 is not fully recruited to the survey/fishery area. Therefore, age 4 has been treated as fully recruited and only the indices from age 4 and higher have been taken into account in the reopening.

## Additional TOR on the historic performance of the reopening protocol

WGNSSK was asked to start with an analysis on the historic performance of the reopening protocol. WGNSSK analysed whether the recruitment assumptions used in June were farer away from assessment estimates in the following years compared to the recruitment assumptions used in autumn incorporating quarter 3 survey information. The recruitment assumptions were compared to assessment estimates in year+1 after the reopening, year+2 and to estimates from the latest assessment. Recruitment estimates may change over time when more information about the cohort becomes available. In addition, benchmarks often lead to changes in assessment results. In total 7 reopening events were analysed for plaice, sole and whiting. The advice for cod and saithe has not been reopened in recent years based on the reopening protocol.

In all cases analysed, the recruitment used for the intermediate year in the October forecasts was closer to the recruitment estimates from assessments in the following years (table 4.1.1). This perception does not change when looking at different assessments (e.g., before and after benchmarks). However, the analysis also revealed the general high uncertainty in recruitment assumptions.

For whiting the 0 group index has been used to update also recruitment assumptions for the TAC year. The analysis revealed that the updated recruitment was farer away from the assessment estimates in the following years compared to the June assumption not using the information from the 0 group index (table 4.2.2). Therefore, the 0 group index seems not to be a reliable proxy for the incoming year class.

Table 4.1.1 Comparison of the performance of intermediate year recruitment assumptions between June and October forecasts

| Intermediate y | cruitment: |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stock | Benchmarks since 2009 | Year of the Reopening | Recruitment for intermediate year used in spring forecast | Recruitment for intermediate year used in autumn forecast | Corresponding recruitment estimate in the assessment carried out in reopening year +1 | Corresponding recruitment estimate in the assessment carried out in reopening year +2 | Correspondi ng <br> recruitment estimate in the 2016 assessment | Absolute percentage difference between recruitment used in spring forecast and recruitment estimate in the assessment carried out in reopening year+1 | Absolute percentage difference between recruitment used in autumn forecast and recruitment estimate in the assessment carried out in reopening year+1 | Absolute percentage difference between recruitment used in spring forecast and recruitment estimate in the assessment carried out in reopening year+2 | Absolute percentage difference between recruitment used in autumn forecast and recruitment estimate in the assessment carried out in reopening year+2 | Absolute percentage difference between recruitment used in spring forecast and recruitment estimate in the 2016 assessment | Absolute <br> percentage <br> difference <br> between <br> recruitment <br> used in autumn <br> forecast and <br> recruitment <br> estimate in the <br> 2016 <br> assessment |
| sol (age2) | 2015 | 2014 | 54268 | 65474 | 78062 | 116775 | 116775 | 30.5 | 16.1 | 53.5 | 43.9 | 53.5 | 43.9 |
| sol (age1) | 2015 | 2015 | 103741 | 135220 | 194127 |  | 194127 | 46.6 | 30.3 | x | $\times$ | 46.6 | 30.3 |
| plaice (age1) | 2015 (merged with 3a) | 2014 | 936981 | 1309243 | 1542295 | 1966051 | 1966051 | 39.2 | 15.1 | 52.3 | 33.4 | 52.3 | 33.4 |
| plaice (age1) | 2015 (merged with 3a) | 2015 | 650882 | 826318 | 1140208 | $\times$ | 1140208 | 42.9 | 27.5 | $\times$ | x | 42.9 | 27.5 |
| WHG47d | 2013 | 2013 | 2139711 | 1119366 | 624701 | 1453422 | 1140360 | 242.52 | 79.18 | 47.22 | 22.98 | 87.63 | 1.84 |
| WHG47d | 2013 | 2014 | 2613817 | 2497236 | 2446622 | 2155672 | 2155672 | 6.83 | 2.07 | 21.25 | 15.84 | 21.25 | 15.84 |
| WHG47d | 2013 | 2015 | 4352809 | 3836431 | 2097757 | x | 2097757 | 107.50 | 82.88 | x | x | 107.50 | 82.88 |

Table 4.1.2 Comparison of the performance of TAC year recruitment assumptions between June and October forecasts


### 4.2 Cod in Subarea 4, 7.d and 3.a

### 4.2.1 New survey information

New survey information, in the form of the IBTS Q3 2016 data, has come to light, subjecting this assessment to the AGCREFA protocol for re-opening advice in the autumn. The Delta-GAM model was re-applied to the full IBTS Q3 time series of North Sea cod data from DATRAS to provide a Q3 index for this stock. The new Delta-GAM Q3 index time series is given in Table 4.2.1.

### 4.2.2 RCT3 analysis

Following the protocol stipulated by AGCREFA (ICES-AGCREFA 2008), an RCT3 analysis was run to provide an estimate of the abundance of the incoming (2015) year-class at age 1. The RCT3 input and output files are given in Tables 4.2.2 and 4.2.3, respectively

### 4.2.3 Update protocol calculations

The outcome of the application of the protocol was as follows:

| CALCULATIONS FOR 2015 YEAR-CLASS AT AGE 1 |  |
| :--- | :---: |
| Log WAP from RCT3 $(R)$ | 11.93 |
| Log of recruitment assumed in spring $(A)$ | 12.19 |
| Int SE of log WAP $(S)$ | 0.158 |
| Distance D $\left(D=\frac{R-A}{S}\right)$ | -1.623 |

### 4.2.4 Conclusions from Protocol

As the distance $\mathrm{D}<-1.0$, the protocol concludes that the advisory process for North Sea cod should be reopened. The autumn indices suggest that the size of the incoming year-class is significantly lower than what had been assumed in the forecast produced by WGNSSK in May 2016.

### 4.2.5 Updated forecast

Given the conclusion of the application of the protocol, the forecast was revised for North Sea cod. The assessment and forecast were re-run with the new Q3 data included, but using the SAM estimate of recruitment for the intermediate year, with recruitment for the years following the intermediate year being resampled, with replacement, from the period 1998 to the final year of catch data. Otherwise the settings and assumptions were unchanged from those used by WGNSSK in May 2016.

Outputs from the assessment re-run with the new Q3 data included are given in Table 4.2.4 and Figure 4.2.1, and the updated catch options Table 4.2.5.

Following the ICES MSY approach, the new short term forecasts lead to a decrease in advised catch from 47431 tonnes to 47359 tonnes (a decrease of 72 tonnes).

## References

ICES-AGCREFA (2008). Report of the Ad hoc Group on Criteria for Reopening Fisheries Advice (AGCREFA). ICES CM 2008/ACOM:60.

Table 4.2.1. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Survey tuning indices for Q3 (NS-IBTS Delta-GAM indices). Data used in the assessment are highlighted in bold font. (The equivalent Q1 index can be found in Section 14, Table 14.6 of this report).


Table 4.2.2. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. RCT3 Inputs.

| "year |  | "recruitment" | "DeltaGAMq11" | "DeltaGAMq31" |
| :---: | :---: | :---: | :---: | :---: |
| 1982 | 946949 | 3741.7822 | NA |  |
| 1983 | 1709993 | 11153.0972 | NA |  |
| 1984 | 415817 | 544.1801 NA |  |  |
| 1985 | 1861699 | 11081.9228 | NA |  |
| 1986 | 709276 | 4343.5409 | NA |  |
| 1987 | 490411 | 2537.0711 | NA |  |
| 1988 | 827364 | 8362.1714 | NA |  |
| 1989 | 327420 | 1770.0654 | NA |  |
| 1990 | 374370 | 1537.8647 | NA |  |
| 1991 | 856834 | 8175.9454 | 16735.7875 |  |
| 1992 | 434521 | 2820.5127 | 4811.7253 |  |
| 1993 | 1018661 | 6187.1008 | 17041.4007 |  |
| 1994 | 596002 | 6769.1817 | 9108.1228 |  |
| 1995 | 372876 | 1639.5325134 .2 |  |  |
| 1996 | 1160081 | 13784.828 | 29167.5344 |  |
| 1997 | 141210 | 1578.9563 | 888.1424 |  |
| 1998 | 251450 | 1305.4656 | 3323.7578 |  |
| 1999 | 457257 | 3481.6615 | 6500.3466 |  |
| 2000 | 167042 | 920.89961406 .5 |  |  |
| 2001 | 246965 | 2474.5739 | 4011.9079 |  |
| 2002 | 123254 | 364.5132983 .71 |  |  |
| 2003 | 201793 | 2486.7931 | 3226.6848 |  |
| 2004 | 154508 | 943.49261109 .1 |  |  |
| 2005 | 358255 | 3493.3633 | 5580.4643 |  |
| 2006 | 168552 | 1202.5336 | 1900.1495 |  |
| 2007 | 196025 | 1910.6205 | 2569.8349 |  |
| 2008 | 193300 | 961.17211983 .08 |  |  |
| 2009 | 296262 | 2508.8754672 .7 |  |  |
| 2010 | 148153 | 662.53831247 .5 |  |  |
| 2011 | 203211 | 1343.4717 | 2127.9124 |  |
| 2012 | 263024 | 1404.7233 | 3246.5997 |  |
| 2013 | 391601 | 2357.5571 | 3498.9504 |  |
| 2014 | 169058 | 1510.2513 | 1906.2281 |  |
| 2015 | NA | 878.72591453 .6 | 21 |  |

Table 4.2.3. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. RCT3 Outputs.

```
Analysis by RCT3 ver4.0
Cod
Data for 2 surveys over 34 years : 1982 - 2015
Regression type = C
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as .00
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.
yearclass:2015
    index slope intercept se rsquare n indices prediction
se.pred WAP.weights
    DeltaGAMq11 0.9635
0.4756 0.115
    DeltaGAMq31 0.7144 6.736 0.1579 0.9447 24 7.282 11.94
0.1715 0.885
        VPA Mean NA NA NA NA 33 12.81
0.7496 0.000
    WAP logWAP int.se
yearclass:2015 151785 11.93 0.1577
```

Table 4.2.4. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Assessment summary. Weights are in tonnes.

| Year |  | TSB <br> (tonnes) Low High | $\begin{array}{rrr} \text { SSB } & & \\ \text { (tonnes) } & \text { Low } & \text { High } \end{array}$ | Fbar 2-4 | Low | High | Landings | Discards | Catch | Unaccounted | Total Removals | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1963 | 487478357668664400 | 512471439164598016 | 152207116938 | 0.473 | 0.410 | 0.546 | 106938 | 10880 | 117830 |  | 117830 | 58 | 3294 |
| 1964 | 802109589683 1091060 | 661986563467 777731 | 163407128333208068 | 0.515 | 0.452 | 0.587 | 135131 | 9818 | 145074 |  | 145074 | 131312 | 160278 |
| 1965 | 10434057695951144632 | 83400977597969306 | 199386 | 0.567 | 0.498 | 45 | 182225 | 17125 | 199187 |  | 199187 | 7 | 2339 |
| 1966 | 1273144 940099 72476 | 997493859127 | 221682 | 0.572 | 0.505 | 0.648 | 214701 | 26318 | 241108 |  | 241108 | 215680 | 269534 |
| 1967 | 10741077923351456082 | 10528389167991209063 | 250446204671306459 | 0.609 | 0.540 | 0.687 | 260928 | 26742 | 287506 |  | 287506 | 25669 | 322012 |
| 1968 | 544161400873738665 | 879404783787986687 | 261450220039310091 | 0.646 | 0.572 | 0.729 | 276509 | 17168 | 293608 |  | 293608 | 266424 | 23565 |
| 1969 | 479740 | 735275650537 | 258074215584 | 0.613 | 0.545 | 0.689 | 217075 | 9685 | 226840 |  | 226840 | 209438 | 245688 |
| 1970 | 15612541499402119689 | 11930229893901438564 | $270222 \begin{array}{llll}226848 & 321890\end{array}$ | 0.650 | 0.581 | 0.727 | 232582 | 19944 | 252458 |  | 252458 | 227946 | 287164 |
| 1971 | 203498714925852774997 | 1330413 124452 1574100 | $274032 \quad 230600325643$ | 0.735 | 0.660 | 0.819 | 291851 | 57931 | 349759 |  | 349759 | 300632 | 40694 |
| 1972 | $508897 \begin{array}{llllll} & 372786 & 694703\end{array}$ | 921723815999104 |  | 0.796 | 0.74 | 887 | 328404 | 34338 | 362580 |  | 362580 | 37387 | 207 |
| 1973 | 7397005421991009292 | $738222654023 \quad 83326$ | 209190 181291 241384 | 0.781 | 0.700 | 0.870 | 234451 | 24884 | 259367 |  | 259367 | 236089 | 284941 |
| 1974 | 725053530396 99149 | 710696 628446 803710 | 227521 197289 262387 | 0.744 | 0.668 | . 830 | 209400 | 26056 | 235390 |  | 235390 | 210316 | 263453 |
| 1975 | $1234282895108 \quad 701975$ | 806130686338 94 | 208147 179111 241890 | 0.802 | 0.722 | . 890 | 208981 | 36062 | 244997 |  | 244997 | 213581 | 281033 |
| 1976 | 845768609101174399 | 636029558569 | 177371 | 0.856 | 0.770 | 0.952 | 201189 | 43695 | 244997 |  | 244997 | 212673 | 82233 |
| 1977 | $2090680 \quad 1514446288667$ | 986580803329121633 | $152512{ }^{129976} \quad 188956$ | 0.815 | 0.733 | 0.906 | 181498 | 77575 | 258849 |  | 258849 | 213081 | ${ }^{14447}$ |
| 1978 | $1329083960376{ }^{1839345}$ | 1125795 937873 1351871 | 15334301855886 | 0.903 | 0.84 | 1.001 | 305896 | 48533 | 354336 |  | 354336 | 291550 | 30642 |
| 1979 | 1634749 1884629 2255901 | 10392408838711221920 | $155282{ }^{188233} 174435$ | 0.846 | 0.764 | 0.938 | 277895 | 61821 | 339762 |  | 339762 | 290618 | 397215 |
| 1980 | 2623448 1892369 3636965 | 1254189 1039136 1513749 | $\begin{array}{llll}171785 & 154018 & 191602\end{array}$ | 0.921 | 0.834 | 1.07 | 290686 | 100509 | 391210 |  | 391210 | 324295 | 471933 |
| 1981 | 105600176634331460691 | 10371638966871199647 | $\begin{array}{lllll}186652 & 168862 & 206315\end{array}$ | 0.937 | 0.85 | 1032 | 342148 | 53745 | 395933 |  | 395933 | 337559 | 464401 |
| 1982 | $1727179{ }^{1265098} 2358037$ | 1132570 947363 1353984 | 1816801163854 | 1.049 | 0.954 | 1.154 | 323191 | 63450 | 386544 |  | 386544 | 327558 | 45615 |
| 1983 | $9450577^{703385}$ | 8855827677521029541 | $153584{ }^{188097} 170807$ | 1.042 | 0.449 | 1.44 | 287794 | 37123 | 324811 |  | 324811 | 2770 | 388847 |
| 1984 | 170999312753872292698 | $908000{ }^{76430108082784}$ | 132058 183313 147440 | 0.974 | 0.887 | 1.070 | 209819 | 67914 | 277895 |  | 277895 | 236097 | ${ }^{327093}$ |
| 1985 |  | 586542518972662910 | $133920{ }^{120029}$ | 0.939 | .854 | 1.033 | 213844 | 27945 | 241832 |  | 241832 | 209545 | 279095 |
| 1986 | 1863562 в92512 2493954 | 818313671822 9967 | 117830 106244 130680 | 0.993 | 0.905 | 1.889 | 168721 | 58924 | 227749 |  | 227749 | 90730 | 271953 |
| 1987 | $709276{ }^{53469} 946570$ | $749379644652 \quad 87119$ | 124244 111709 138184 | 0.976 | 890 | 1072 | 225258 | 32598 | 257816 |  | 257816 | 217746 | 305259 |
| 1988 | 490411 367096 655150 | 550730481290630188 | 122394 11193 134000 | 0.997 | 0.909 | 1.093 | 191377 | 14698 | 206076 |  | 206076 | 182779 | 32342 |
| 1989 | 8298506187771112923 | 556265470358 6578 | 109754 | 1.014 | 0.924 | 113 | 138968 | 40336 | 179333 |  | 179333 | 5 | 20899 |
| 1990 | 327093245507435792 | 371759327430422088 | 9931089720 | 0.940 | 0.853 | 1.035 | 115151 | 23063 | 138275 |  | 138275 | 120936 | 158.00 |
| 1991 | 374370282733495707 | 342491298970393346 | 95511885782106344 | 0.931 | 0.846 | 1.024 | 102437 | 15709 | 118184 |  | 118184 | 104964 | ${ }^{133070}$ |
| 1992 | $861991655465 \quad 1133590$ | 537132448600 | 91309 | 0.921 | 838 | 1.012 | 108662 | 31477 | 140225 |  | 140225 | 18377 | 16604 |
| 1993 | $435827 \begin{array}{lllll} & 334516 & 567820\end{array}$ | 415817 365903 47253 | 98913889968108748 | 0.940 | 0.855 | 1.033 | 130180 | 28557 | 158843 | 9947 | 148896 | 128471 | 7256 |
| 1994 | $1022744{ }^{777610} 1845155$ | 531256449397628027 | 101926 93458 $\quad 11160$ | 0.963 | 0.879 | 1056 | 106186 | 41938 | 148146 | 5899 | 154045 | 132790 | 78703 |
| 1995 | 598990457760 | 565802488675655103 | 121419 111338 | 1.010 | 0.922 | 1.107 | 130522 | 31962 | 162478 | 2857 | 190995 | 163662 | 22889 |
| 1996 | $374745 \quad 287478488503$ | 422101373015477646 | 116658 107469 | 1.011 | 0.923 | 1.107 | 132366 | 21522 | 153941 | 2276 | 156217 | 38561 | 7662 |
| 1997 | 117174088819761556702 | 647582529009792732 | $101417{ }^{93291}$ | 0.988 | 0.903 | 1.081 | 133460 | 46610 | 180153 | 2580 | 154353 | 29050 | 84619 |
| 1998 | 141351 107469 18594 | $329062 \quad 289708373760$ | $102847{ }^{93115}$ | 1.007 | 0.921 | 1.100 | 148023 | 43876 | 191975 | 55622 | 136353 | 16955 | 968 |
| 1999 | 252458 194125 3288319 | 227067203490 | 85648 | 1.064 | 0.974 | 1.163 | 96833 | 13896 | 110718 | ${ }^{15683}$ | 95035 | 12 | 104037 |
| 2000 | 459089 | 290106247415 | $68118 \quad 61981 \quad 74863$ | 1.073 | 981 | 1.73 | 73539 | 16615 | 90091 | -4785 | 85306 | 73852 | 98536 |
| 2001 | $166542 \quad 17920216884$ | 198392 75980 223659 | 63386 | 1.005 | 0.915 | 102 | 44499 | 11451 | 55950 | 16453 | 72403 | 63868 | 82078 |
| 2002 | 249447 199925 324208 | 169228 148815 192440 | $56050 \quad 50967 \quad 61640$ | 0.951 | 0.865 | ${ }^{1046}$ | 53494 | 11512 | 65012 | 399 | 56613 | 5180 | 623 |
| 2003 | 122516 93928 1159807 | $\begin{array}{llll}141917 & 27458 & 15807\end{array}$ | $56444{ }^{51365} \quad 62025$ | 0.927 | 0.838 | 1.226 | 31280 | 4788 | 36088 | 82 | 53370 | 47856 | 5959 |
| 2004 | 202805156392262992 | $\begin{array}{llll}123871 & 108727 & 14125\end{array}$ | $45844 \begin{array}{llll}4235 & 50968\end{array}$ | 0.890 | 0.803 | 0.987 | 27316 | 7546 | 34865 | 454 | 39379 | 35829 | 43282 |
| 2005 | 154353 17795 202258 | 138690 ²207 158696 | 47335 4742 53677 | 0.828 | 0.744 | 0.920 | 29923 | 11382 | 41291 | -236 | 40055 | 35443 | 45266 |
| 2006 | 359331277468465348 | 146679 23851 | $43002 \quad 37440 \quad 49390$ | 0.734 | 0.654 | 0.824 | 22652 | 9136 | 31793 |  | 31793 | 28245 | 35787 |
| 2007 | $168721{ }_{130724} 217762$ | 194853 п2032 220701 | 72475 6400 | 0.678 | 0.602 | 0.764 | 24029 | 29144 | 53157 |  | 53157 | 46548 | 60704 |
| 2008 | 197402 152732 255688 | 205870 180652 234608 | $80822 \quad 74493 \quad 91368$ | 0.643 | 0.566 | 0.730 | 27065 | 25311 | 52365 |  | 52365 | 47603 | 5764 |
| 2009 | $193300 \begin{array}{llll}149564 & 249826\end{array}$ |  | $9049079262 \quad 103309$ | 0.629 | 0.551 | 0.719 | 33290 | 21673 | 54940 |  | 54940 | 49673 | 60766 |
| 2010 | 297450229280385887 | 235861203787272984 | 92411 79388 107570 | 0.544 | 771 | 0.629 | 36207 | 12565 | 48776 |  | 48776 | 44312 | 53691 |
| 2011 | $148153{ }^{143375} \quad 19906$ | 223463 ต3941 257479 | $10492587982 \quad$ 12581 | 0.444 | . 380 | 518 | 34372 | 10446 | 44846 |  | 44846 | 40490 | 49672 |
| 2012 | 203618157642263003 | 198988 171971 230249 | 10608587895127958 | 0.403 | 0.344 | 0.472 | 32696 | 7618 | 40336 |  | 40336 | 37289 | 43632 |
| 2013 | 267533206741346203 | 260928224014303924 | 116891 97007 140852 | 0.391 | 0.336 | 0.455 | 30792 | 10777 | 41564 |  | 41564 | 38186 | 524 |
| 2014 | $3991133^{302761} 526128$ | $333701282322 \quad 394430$ | 127262 10649 ${ }^{152573}$ | 0.389 | 35 | 451 | 34787 | 11086 | 45844 |  | 45844 | 4567 | 51 |
| 2015 | 1749051274145 | $294785 \quad 252701343877$ | $153584{ }^{127598} 8184861$ | 0.371 | 0.316 | 0.434 | 38561 | 13558 | 52104 |  | 52104 | 46732 | 58094 |
| 2016 | 134054 |  |  |  |  |  |  |  |  |  |  |  |  |

Table 4.2.5. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Catch options. Units are ‘000t (SSB landings, discards, unaccounted) or millions (recruitment).

| Intermediate year F assumption: $\mathrm{F}(2016)=\mathrm{F}(2015)=0.37$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Recruitment resam SSB(2017) = HC landings (2016) Discards (2016) = | led from | 1998-2015 = |  | 197 <br> 176299 <br> 46031 <br> 10930 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Rationale | $\begin{aligned} & \text { Catch } \\ & \text { (2017) } \\ & \hline \end{aligned}$ | Landings (2017) | $\begin{gathered} \hline \text { Discards } \\ (2017) \\ \hline \end{gathered}$ | Basis | $\begin{aligned} & \hline \text { Ftotal } \\ & (2017) \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \text { F land } \\ & (2017) \end{aligned}$ | $\begin{gathered} \hline \text { F disc } \\ (2017) \\ \hline \end{gathered}$ | SSB (2018) | $\begin{gathered} \hline \text { SSB 5\% } \\ (2018) \\ \hline \end{gathered}$ | $\begin{array}{r} \% \text { SSB } \\ \text { change } \end{array}$ | $\begin{array}{r} \hline \% \mathrm{TAC} \\ \text { change } \end{array}$ | $\begin{aligned} & \hline \text { Ftotal } \\ & (2018) \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \text { Ftotal } \\ & (2019) \end{aligned}$ | $\begin{array}{r} \text { Catch } \\ (2018) \\ \hline \end{array}$ | $\begin{aligned} & \text { Catch } \\ & (2019) \\ & \hline \end{aligned}$ | Landings <br> (2018) | Landings <br> (2019) | Discards <br> (2018) | Discards (2019) | $\begin{array}{r} \text { SSB } \\ (2019) \\ \hline \end{array}$ | $\begin{array}{r} \text { SSB \%change \%change } \\ (2020) \text { SSB 19:17 SSB 20:17 } \end{array}$ |  |  |
| Management Plan | 55876 | 46754 | 9122 | EU MP | 0.40 | 0.28 | 0.12 | 171971 | 137734 | -2 | 16 | 0.40 | 0.40 | 51685 | 48948 | 41834 | 38355 | 9851 | 10593 | 168267 | 162985 | -5 | -8 |
| Management Plan | 55876 | 46754 | 9122 | EU-Norway | 0.40 | 0.28 | 0.12 | 171971 | 137734 | -2 | 16 | 0.40 | 0.40 | 51685 | 48948 | 41834 | 38355 | 9851 | 10593 | 168267 | 162985 | -5 | -8 |
| MSY approach | 47359 | 39651 | 7708 | FMSY *SSB2017/Btrigger | 0.33 | 0.23 | 0.10 | 181374 | 146059 | 3 | -2 | 0.33 | 0.33 | 46112 | 45035 | 37545 | 35727 | 8567 | 9308 | 185027 | 185341 | 5 | 5 |
| Zero Catch | 0 | 0 | 0 | F=0 | 0.00 | 0.00 | 0.00 | 236794 | 192601 | 34 | -100 | 0.00 | 0.00 | 0 | 0 | 0 | 0 | 0 | 0 | 299892 | 356630 | 70 | 102 |
| MSY | 47359 | 39651 | 7708 | FMSY | 0.33 | 0.23 | 0.10 | 181374 | 146059 | 3 | -2 | 0.33 | 0.33 | 46112 | 45035 | 37545 | 35727 | 8567 | 9308 | 185027 | 185341 | 5 | 5 |
| Fpa | 57039 | 47740 | 9299 | Flim/1.4 | 0.41 | 0.29 | 0.12 | 170652 | 136590 | -3 | 18 | 0.41 | 0.41 | 52391 | 49363 | 42409 | 38673 | 9982 | 10690 | 166166 | 160137 | -6 | -9 |
| Flim | 75481 | 63192 | 12289 | Flim | 0.58 | 0.41 | 0.17 | 149955 | 118849 | -15 | 56 | 0.58 | 0.58 | 61568 | 54295 | 49196 | 41415 | 12372 | 12880 | 132551 | 120002 | -25 | -32 |
| SSB(2018)=Blim | 105033 | 87793 | 17240 | SSB(2018)=Blim | 0.91 | 0.64 | 0.27 | 118000 | 91689 | -33 | 117 | 0.91 | 0.91 | 69219 | 55548 | 53523 | 39560 | 15696 | 15988 | 88774 | 73717 | -50 | -58 |
| SSB(2018)=Bpa | 61928 | 51863 | 10065 | SSB(2018)=Bpa | 0.45 | 0.32 | 0.13 | 165000 | 131881 | -6 | 28 | 0.45 | 0.45 | 55208 | 51159 | 44551 | 39696 | 10657 | 11463 | 156600 | 148228 | -11 | -16 |
| SSB(2018)=Btrigger | 61928 | 51863 | 10065 | SSB(2018)=Btrigger | 0.45 | 0.32 | 0.13 | 165000 | 131881 | -6 | 28 | 0.45 | 0.45 | 55208 | 51159 | 44551 | 39696 | 10657 | 11463 | 156600 | 148228 | -11 | -16 |
| TAC constraint | 38404 | 32335 | 6069 | TAC2016-20\% | 0.26 | 0.18 | 0.08 | 192162 | 151192 | 9 | -20 | 0.26 | 0.26 | 39313 | 40286 | 32335 | 32335 | 6978 | 7951 | 204463 | 211131 | 16 | 20 |
| TAC constraint | 40813 | 34356 | 6457 | TAC2016-15\% | 0.28 | 0.20 | 0.08 | 189519 | 148841 | 7 | -15 | 0.28 | 0.29 | 41844 | 42950 | 34356 | 34356 | 7488 | 8594 | 198587 | 202280 | 13 | 15 |
| TAC constraint | 43224 | 36377 | 6847 | TAC2016-10\% | 0.30 | 0.21 | 0.09 | 186772 | 146198 | 6 | -10 | 0.31 | 0.32 | 44394 | 45633 | 36377 | 36377 | 8017 | 9256 | 192462 | 192448 | 9 | 9 |
| TAC constraint | 45635 | 38398 | 7237 | TAC2016-5\% | 0.32 | 0.22 | 0.10 | 184025 | 143430 | 4 | -5 | 0.33 | 0.35 | 46951 | 48338 | 38398 | 38398 | 8553 | 9940 | 186436 | 183751 | 6 | 4 |
| TAC constraint | 48049 | 40419 | 7630 | TAC2016 | 0.33 | 0.24 | 0.09 | 181234 | 140614 | 3 | 0 | 0.36 | 0.39 | 49517 | 51069 | 40419 | 40419 | 9098 | 10650 | 180418 | 174460 | 2 | -1 |
| TAC constraint | 50464 | 42440 | 8024 | TAC2016+5\% | 0.35 | 0.25 | 0.10 | 178427 | 138147 | 1 | 5 | 0.39 | 0.43 | 52094 | 53858 | 42440 | 42440 | 9654 | 11418 | 174548 | 165199 | -1 | -6 |
| TAC constraint | 52879 | 44461 | 8418 | TAC2016 + 10\% | 0.37 | 0.26 | 0.11 | 175714 | 135682 | 0 | 10 | 0.42 | 0.47 | 54685 | 56669 | 44461 | 44461 | 10224 | 12208 | 168533 | 156066 | -4 | $-11$ |
| TAC constraint | 55295 | 46482 | 8813 | TAC2016 + 15\% | 0.39 | 0.28 | 0.11 | 173027 | 133064 | -2 | 15 | 0.45 | 0.52 | 57281 | 59528 | 46482 | 46482 | 10799 | 13046 | 162639 | 146886 | -8 | -17 |
| TAC constraint | 57713 | 48503 | 9210 | TAC2016 + 20\% | 0.42 | 0.29 | 0.13 | 170353 | 130529 | -3 | 20 | 0.48 | 0.57 | 59906 | 62430 | 48503 | 48503 | 11403 | 13927 | 156576 | 137768 | -11 | -22 |
| Status quo | 52735 | 44156 | 8579 | Fsq | 0.37 | 0.26 | 0.11 | 175461 | 140750 | 0 | 9 | 0.37 | 0.37 | 49727 | 47606 | 40346 | 37455 | 9381 | 10151 | 174146 | 170958 | -1 | -3 |



Figure 4.2.4. Cod in Subarea 4 and Divisions 3.a (Skagerrak) and 7.d. Summary of stock assessment with point-wise $95 \%$ confidence intervals. The SAM assessment produced by WGNSSK in May 2016 is plotted in grey for comparison.

### 4.3 Haddock in Sub-Area IV and Divisions IIIa and Via

The reopening protocol has not been applied because the final assessment carried out in October already includes the 3rq quarter survey information from 2016.

### 4.4 Saithe in Subarea 4, 6 and Division 3a

### 4.4.1 New survey information

New survey data are available from the 2016 international third quarter IBTS survey (IBTS Q3) for a potential autumn forecast. The following analysis compares the effect of the new survey data with the forecast provided by the relevant assessment Working Group (ICES-WGNSSK 2016), according to the protocol specified by the ICES Ad hoc Group on Criteria for Reopening Fisheries Advice (ICES-AGCREFA 2008).

### 4.4.2 RCT3 analysis

An RCT3 analysis, following the protocol outlined by AGCREFA (ICES-AGCREFA 2008), was run to provide an estimate of the abundance of the incoming age 3 and age 4 year-classes. The RCT3 input and output files are given in Tables 4.4.1 to 4.4.3.

### 4.4.3 Update protocol calculation

The outcome of following the protocol was:

| CALCuLATion of 2013 Year-CLASS AT: | AGe 3 | AGe 4 |
| :--- | :--- | :--- |
| Log WAP from RCT3 $(\mathrm{R})$ | 11.94 | 11.55 |
| WAP | 153277 | 103777 |
| Log of recruitment assumed in spring $(\mathrm{A})$ | 11.53 | 11.29 |
| Int SE of log WAP $(\mathrm{S})$ | 0.24 | 0.26 |
| Distance $\mathrm{D}^{=}(R-A) / S$ | 1.71 | 1.00 |

### 4.4.4 Conclusions from protocol

As the distance $\mathrm{D}>1.71$, the protocol concludes that the advisory process for North Sea saithe should be reopened. The autumn indices suggest that the size of the incoming year-class is considerably larger than the median value assumed in the forecast produced by WGNSSK in May 2016. However, caution is warranted because the internal consistency of the Q3 survey between age 3 and age 4 were very poor (correlation $=$ 0.24 ; Figure 4.4.1). This indicates that the age 3 index value is very uncertain and the perception of the year class strength can change considerably in the following year when information at age 4 becomes available. Age 3 is only partially recruited to the survey/fishing area, which explaines the low internal consistency to some extent. Age 4 can be seen as first fully recruited year class although the assessment starts with age 3. Therefore, an RCT3 analysis for age 4 was also explored and $D=1$, indicating that age 4 is also marginally larger than expected. Therefore, just following the protocol would imply a reopening of the advice.

### 4.4.5 Updated forecast

The assessment was revised for North Sea saithe, incorporating the new survey data for 2016. The forecast was updated, but the 2016 recruitment estimate from the IBTS Q3 2016 survey was not used in the forecast for the reasons stated above. Instead, recruitment in 2016 was resampled from the period between 2003 and 2015. The settings and assumptions for the forecast are in Table 4.4.4.
Table 4.4.5 gives the draft advice table for the October 2016 update. On this basis, predicted North Sea total catch in 2017 increases from 116605 t (June Advice) to 134962 t (October results), while the corresponding TAC change increases from $62 \%$ to $88 \%$. The very optimistic forecast that results from this reopening has very high uncertainties associated with it (Figure 4.4.2), which are largely ignored in the way the current advice is set up. A $62 \%$ TAC increase has been advised in June together with additional consideration that TAC constraints should be taken into account given the high uncertainty in the assessment and forecast. An even higher increase in the advised TAC would contradict the TAC constraint statement given that the former saithe EU- Norway management strategy had a $15 \%$ TAC constraint.

## References

ICES-AGCREFA (2008). Report of the Ad hoc Group on Criteria for Reopening Fisheries Advice (AGCREFA). ICES CM 2008/ACOM:60.
ICES-WGNSSK (2016). Report of the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), 26 April-5 May 2016, Hamburg, Germany. ICES CM 2016/ ACOM:14. 1023 pp.

Table 4.4.1. Saithe in Subareas 4 and 6 and Division 3a. RCT3 data input file for the age 3 and age 4 year-classes.

| Year class | Age3 |  | Age 4 |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Recruitment | IBTS Q3 | Recruitment | IBTS Q3 |
| 1987 | - | - | 71728 | 0.402 |
| 1988 | 174125 | 1.946 | 88437 | 2.76 |
| 1989 | 103944 | 1.077 | 59297 | 2.781 |
| 1990 | 176170 | 7.965 | 97290 | 1.615 |
| 1991 | 118110 | 1.117 | 66737 | 2.501 |
| 1992 | 215385 | 13.959 | 148907 | 6.533 |
| 1993 | 119594 | 3.825 | 79763 | 3.351 |
| 1994 | 148952 | 3.756 | 120609 | 4.134 |
| 1995 | 87723 | 1.181 | 55318 | 1.907 |
| 1996 | 111435 | 2.086 | 93077 | 8.836 |
| 1997 | 97569 | 3.479 | 67645 | 6.173 |
| 1998 | 205817 | 21.494 | 144410 | 18.974 |
| 1999 | 163051 | 10.748 | 123163 | 23.802 |
| 2000 | 166920 | 19.272 | 109637 | 6.727 |
| 2001 | 117267 | 4.93 | 75797 | 7.512 |
| 2002 | 143615 | 8.916 | 123653 | 29.579 |
| 2003 | 102098 | 10.553 | 55604 | 5.578 |
| 2004 | 155694 | 34.006 | 98547 | 5.584 |
| 2005 | 73631 | 3.312 | 51468 | 1.703 |
| 2006 | 58640 | 1.346 | 38225 | 0.964 |
| 2007 | 90077 | 1.361 | 79213 | 8.451 |
| 2008 | 83142 | 4.52 | 48196 | 2.497 |
| 2009 | 141320 | 11.134 | 107991 | 16.279 |
| 2010 | 99244 | 14.701 | 74017 | 3.923 |
| 2011 | 60074 | 1.649 | 44478 | 5.613 |
| 2012 | 107275 | 11.001 | NA | 17.307 |
| 2013 | NA | 37.74 | - | - |

Table 4.4.2. Saithe in Subareas 4 and 6 and Division 3a. RCT3 data output file for the age 3 yearclass.

```
Analysis by RCT3 R ver3.1 of data from file:
RCT3 Saithe AGE 3 2016.txt
RCT3 input for D calculations for Saithe in Subareas IV VI and Division IIIa
Data for 1 surveys over 26 years: 1988 - 2013
Regression type = c
Tapered time weighting applied
Power = 3 over 20 years
Survey weighting not applied
Final estimates shrunk towards mean
Minimum S.E. for any survey taken as 0.000
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.
yearclass = 2013
                I-----------Regression----------I --------------Prediction-----------
I
```



Table 4.4.3. Saithe in Subareas 4 and 6 and Division 3a. RCT3 data output file for the age 4 yearclass.

```
Analysis by RCT3_R ver3.1 of data from file :
RCT3 Saithe AGE 4 2016.txt
RCT3 input for D calculations for Saithe in Subareas IV VI and Division
IIIa
Data for 1 surveys over 26 years : 1987-2012
Regression type = c
Tapered time weighting applied
Power = 3 over 20 years
Survey weighting not applied
Final estimates shrunk towards mean
Minimum S.E. for any survey taken as 0.000
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.
yearclass = 2012
I------------Regression--------------
--I
Survey/ Slope Inter- Std Rsquare No. Index Predicted Std WAP
Series cept Error Pts Value Value Error Weights
    IBTSQ3 0.62 9.97 0.28 0.704 25 2.91 
\begin{tabular}{lclccccr} 
Year & Weighted & Log & Int & Ext & Var & VPA & Log \\
Class & Average & WAP & Std & Std & Ratio & VPA \\
& Prediction & & Error & Error & & \\
2012 & 103538 & 11.55 & 0.26 & 0.29 & 1.20 &
\end{tabular}
```

Table 4.4.4. Saithe in Subareas 4 and 6 and Division 3a. The basis for the catch options.

| Variable | VaLue | Notes |
| :--- | :--- | :--- |
| F ages 4-7 (2016) | F $=0.20$ | TAC constraint (68601 tonnes)* |
| SSB (2016) | 275345 t | SSB in the intermediate year, tonnes |
| SSB (2017) | 337973 t | SSB at the beginning of the TAC year, tonnes |
| Rage3 (2016) | 109 <br> million | Median recruitment re-sampled from the years 2003-2015 |
| Rage3 (2017) | 109 <br> million | Median recruitment re-sampled from the years 2003-2015 |
| Total catch (2016) | 72335 t | Assuming 2015 landings fraction by age, tonnes |
| Commercial landings <br> (2016) | 68601 t | TAC 2015, tonnes |
| Discards (2016) | 3734 t | Assuming 2015 discard fraction by age, tonnes |

* TAC was based on landings without adjustment.

Table 4.4.5. Saithe in Subareas 4 and 6 and Division 3a. Draft advice table; all weights are in tonnes.

| Rationale | Total CATCH (2017) | Wanted CATCH* (2017) | UnWANTED <br> CATCH* <br> (2017) | Wanted CATCH 3.A \& 4 (2017) ** | Wanted $\begin{gathered} \text { CATCH } \\ 6 \\ (2017) \end{gathered}$ | BASIS | Ftotal (2017) | Fwanted (2017) | Funwanted (2017) | $\begin{aligned} & \text { SSB } \\ & (2018) \end{aligned}$ | $\begin{gathered} \% \text { SSB } \\ \text { CHANGE } \\ * * * \end{gathered}$ | \% TAC <br> CHANGE <br> WANTED <br> CATCH^ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MSY <br> approach | 140653 | 134792 | 5861 | 122122 | 12670 | FMSY | 0.36 | 0.34 | 0.02 | 333297 | -1 | 96 |
| EU-Norway management strategy | 82455 | 78976 | 3479 | 71552 | 7424 | Paragraph 5 of management strategy | 0.2 | 0.19 | 0.01 | 390772 | 16 | 15 |
| Zero catch | 0 | 0 | 0 | 0 | 0 | $\mathrm{F}=0$ | 0 | 0 | 0 | 470855 | 39 | -100 |
| Other options | 83984 | 80439 | 3544 | 72878 | 7561 | F2016 | 0.2 | 0.19 | 0.01 | 389271 | 15 | 17 |
|  | 71912 | 68601 | 3311 | 62153 | 6448 | TAC2016 | 0.17 | 0.16 | 0.01 | 400429 | 18 | 0 |
|  | 152927 | 146546 | 6381 | 132771 | 13775 | Fpa | 0.4 | 0.38 | 0.02 | 321560 | -5 | 114 |
|  | 201882 | 193230 | 8651 | 175066 | 18164 | Flim | 0.56 | 0.54 | 0.02 | 273675 | -19 | 182 |
|  | 385729 | 365782 | 19946 | 331398 | 34384 | $\begin{aligned} & \text { SSB2018 = } \\ & \text { Blim } \end{aligned}$ | 1.58 | 1.5 | 0.08 | 107000 | -68 | 433 |
|  | 335831 | 319572 | 16259 | 289532 | 30040 | $\begin{aligned} & \text { SSB2018= } \\ & \text { Bpa } \end{aligned}$ | 1.2 | 1.15 | 0.05 | 150000 | -56 | 366 |
|  | 335831 | 319572 | 16259 | 289532 | 30040 | $\begin{aligned} & \text { SSB2018 = } \\ & \text { MSY } \\ & \text { Btrigger } \end{aligned}$ | 1.2 | 1.15 | 0.05 | 150000 | -56 | 366 |

[^9]${ }^{\wedge}$ Wanted catch 2017 relative to the 2016 wanted catch (without adjustment) TAC.


Figure 4.4.1. Saithe in Subareas 4 and 6 and Division 3a. Internal consistencies between subsequent ages in the IBTS Q3 survey.


Figure 4.4.2. Saithe in Subareas 4 and 6 and Division 3a. Historic trend and three-year projection in $F_{4.7}$ SSB, recruitment, and catch including the associated uncertainties in the estimates.

### 4.5 Whiting in Sub-Area IV and Division VIId

### 4.5.1 New survey information

The new data available for a potential autumn forecast are the international third-quarter North Sea IBTS survey (IBTS Q3). The full available dataset for the IBTS Q3 series is given in Table 4.5.1. Note that the following analysis compares the effect of the new survey data with the forecast provided by the relevant assessment Working Group (ICES-WGNSSK 2016), according to the protocol specified by the ICES Ad hoc Group on Criteria for Reopening Fisheries Advice (ICES-AGCREFA 2008).

### 4.5.2 RCT3 analysis

Following the protocol stipulated by AGCREFA (ICES-AGCREFA 2008), an RCT3 analysis was run to provide an estimate of the abundance of the 2015 year-class at age 0 , and the incoming (2016) year-class at age 0 . The RCT3 input and output files are given in Tables 4.5.2 and 4.5.3.

### 4.5.3 Update protocol calculations

The outcome of the application of the protocol was as follows:

| CALCULATIONS FOR 2015 YeAr-CLASS AT AGE 1 (IN 2016) |  |
| :--- | :--- |
| Log WAP from RCT3 $(R)$ | 14.73 |
| Log of recruitment assumed in spring $(A)$ | 14.88 |
| Int SE of log WAP $(S)$ | 0.154 |
| Distance D $\left(D=\frac{R-A}{S}\right)$ | -0.97 |


| CALCULATIONS FOR 2016 YEAR-CLASS AT AGE 1 (IN 2017) |  |
| :--- | :--- |
| Log WAP from RCT3 $(R)$ | 14.92 |
| Log of recruitment assumed in spring $(A)$ | 14.71 |
| Int SE of log WAP $(S)$ | 0.284 |
| $\left(D=\frac{R-A}{S}\right)$ | 0.74 |
| Distance D |  |

### 4.5.4 Conclusions from protocol

- 2015 year class at age 1 (in 2016): as the distance $-1.0<\mathrm{D}<1.0$, the protocol concludes that the RCT3 estimate used for 2015 recruitment was appropriate and need not be changed.
- 2016 year class at age 1 (in 2017): in the spring advice, a geometric mean value was used for this year class. As the distance $-1.0<\mathrm{D}<1.0$ for this year-class, the protocol concludes that the original geometric mean appropriate and need not be changed.

The overall conclusion is that the advisory process for North Sea whiting does not have to be reopened based on the RCT3 analysis.

### 4.5.5 Reopening performance

Since 2009, the advice had to be reopened and updated three times (in 2013, 2014 and 2015). Generally the autumn update for the intermediate year improved the forecast from spring relative to the values estimated in the following years (reopening year+1, reopening year +2 or 2016). The percentage difference relative to 2016 could be reduced from 88, 21 and $108 \%$ in spring to 2,16 and $83 \%$ in autumn, respectively.

In contrast for the TAC year, the spring forecast using the geometric mean of recruitment of recent years was better in forecasting the recruitment estimated in the following years than the autumn update. In 2014 and 2015 when recruitment for the TAC year was updated, the percentage difference to the respective estimate in the final year 2016 increased from 76 and $30 \%$ to 188 and $179 \%$, respectively.

### 4.5.6 References

ICES-AGCREFA (2008). Report of the Ad hoc Group on Criteria for Reopening Fisheries Advice (AGCREFA). ICES CM 2008/ACOM:60.

ICES-WGNSSK (2016). Report of the Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak. ICES CM 2015/ACOM:13.

Table 4.5.1. Whiting in Sub-Area IV and Division VIId. Indices from the third-quarter IBTS (IBTS Q3) groundfish survey series. New data from autumn 2016 are highlighted.

| Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 |
| 1991 | 536.99 | 703.368 | 158.594 | 79.024 | 14.568 | 5.183 | 1.018 |
| 1992 | 1379.459 | 600.867 | 296.1 | 72.451 | 57.498 | 10.273 | 6.212 |
| 1993 | 919.193 | 638.722 | 177.377 | 66.118 | 14.711 | 15.904 | 3.039 |
| 1994 | 610.743 | 677.645 | 219.541 | 74.71 | 19.506 | 4.722 | 3.16 |
| 1995 | 729.246 | 619.786 | 291.18 | 107.195 | 21.512 | 6.013 | 3.464 |
| 1996 | 316.501 | 545.708 | 278.218 | 129.356 | 34.003 | 6.893 | 4.1 |
| 1997 | 2062.67 | 332.968 | 180.681 | 108.985 | 28.006 | 10.711 | 4.245 |
| 1998 | 2631.69 | 330.6 | 150.205 | 52.766 | 31.01 | 11.179 | 4.695 |
| 1999 | 2498.55 | 1203.501 | 190.643 | 53.932 | 24.452 | 9.529 | 4.179 |
| 2000 | 1961.467 | 940.784 | 326.515 | 64.396 | 13.597 | 6.534 | 4.861 |
| 2001 | 3548.815 | 668.907 | 283.081 | 93.978 | 19.076 | 4.279 | 6.023 |
| 2002 | 269.285 | 811.915 | 257.157 | 131.47 | 35.034 | 5.45 | 2.835 |
| 2003 | 356.523 | 257.637 | 292.805 | 128.67 | 67.944 | 17.313 | 4.767 |
| 2004 | 714.27 | 150.623 | 59.032 | 66.326 | 45.724 | 27.103 | 9.711 |
| 2005 | 169.321 | 171.386 | 68.259 | 31.433 | 45.616 | 33.96 | 28.704 |
| 2006 | 198.949 | 174.625 | 86.336 | 32.619 | 13.511 | 23.287 | 25.714 |
| 2007 | 822.902 | 95.495 | 63.592 | 37.636 | 11.482 | 8.405 | 20.747 |
| 2008 | 764.759 | 362.299 | 68.886 | 30.907 | 13.774 | 4.081 | 14.791 |
| 2009 | 593.801 | 585.529 | 384.777 | 40.984 | 12.295 | 8.037 | 6.808 |
| 2010 | 510.123 | 224.321 | 145.671 | 54.635 | 12.844 | 5.996 | 7.795 |
| 2011 | 247.085 | 446.812 | 144.439 | 47.243 | 16.217 | 6.929 | 4.635 |
| 2012 | 306.812 | 256.718 | 193.523 | 57.001 | 20.081 | 10.644 | 5.384 |
| 2013 | 334.257 | 67.451 | 60.102 | 65.787 | 17.504 | 7.08 | 3.725 |
| 2014 | 1401.008 | 223.4 | 97.962 | 65.552 | 33.278 | 10.311 | 6.849 |
| 2015 | 2091.636 | 312.453 | 222.551 | 43.072 | 24.038 | 18.433 | 10.853 |
| 2016 | 971.324 | 297.483 | 243.642 | 77.638 | 12.211 | 8.053 | 9.947 |

Table 4.5.2 Whiting in Sub-Area IV and Division VIId. RCT3 input file. Data from surveys in autumn 2016 are highlighted.

| Year <br> class | VPA Recruits at age 1 | IBTS Q1 Age 1 | IBTS Q1 Age 2 | $\begin{gathered} \text { IBTS Q3 } \\ \text { Age } 0 \end{gathered}$ | IBTS Q3 Age 1 | $\begin{gathered} \text { IBTS Q3 } \\ \text { Age } 2 \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 3602988 | 518.936 | 686.445 | -11 | -11 | 158.594 |
| 1990 | 3561689 | 1007.621 | 665.714 | -11 | 703.368 | 296.1 |
| 1991 | 3429482 | 907.297 | 522.811 | 536.99 | 600.867 | 177.377 |
| 1992 | 3911169 | 1075.624 | 627.406 | 1379.459 | 638.722 | 219.541 |
| 1993 | 3665192 | 721.709 | 448.484 | 919.193 | 677.645 | 291.18 |
| 1994 | 3286257 | 678.59 | 485.968 | 610.743 | 619.786 | 278.218 |
| 1995 | 2338049 | 502.361 | 342.246 | 729.246 | 545.708 | 180.681 |
| 1996 | 1785364 | 287.779 | 160.695 | 316.501 | 332.968 | 150.205 |
| 1997 | 2473817 | 543.117 | 305.445 | 2062.67 | 330.6 | 190.643 |
| 1998 | 4023722 | 676.266 | 544.86 | 2631.69 | 1203.501 | 326.515 |
| 1999 | 4784856 | 767.887 | 592.395 | 2498.55 | 940.784 | 283.081 |
| 2000 | 3996189 | 614.174 | 342.774 | 1961.467 | 668.907 | 257.157 |
| 2001 | 3432777 | 558.505 | 298.408 | 3548.815 | 811.915 | 292.805 |
| 2002 | 1210399 | 131.588 | 90.134 | 269.285 | 257.637 | 59.032 |
| 2003 | 1225220 | 184.399 | 97.824 | 356.523 | 150.623 | 68.259 |
| 2004 | 1599745 | 112.663 | 125.057 | 714.27 | 171.386 | 86.336 |
| 2005 | 1600655 | 184.411 | 147.304 | 169.321 | 174.625 | 63.592 |
| 2006 | 1526369 | 64.53 | 205.798 | 198.949 | 95.495 | 68.886 |
| 2007 | 2767240 | 268.598 | 295.812 | 822.902 | 362.299 | 384.777 |
| 2008 | 2263468 | 211.202 | 224.795 | 764.759 | 585.529 | 145.671 |
| 2009 | 2335777 | 326.192 | 337.096 | 593.801 | 224.321 | 144.439 |
| 2010 | 3053030 | 184.867 | 588.309 | 510.123 | 446.812 | 193.523 |
| 2011 | 1612011 | 231.255 | 162.985 | 247.085 | 256.718 | 60.102 |
| 2012 | 1140360 | 54.431 | 184.517 | 306.812 | 67.451 | 97.962 |
| 2013 | 2155672 | 265.226 | 212.642 | 334.257 | 223.4 | 222.551 |
| 2014 | -11 | 315.019 | 312.194 | 1401.008 | 312.453 | 243.642 |
| 2015 | -11 | 281.272 | -11 | 2091.636 | 297.483 | -11 |
| 2016 | -11 | -11 | -11 | 971.324 | -11 | -11 |

Table 4.5.3. Whiting in Sub-Area IV and Division VIId. RCT3 output file.
Analysis by RCT3 ver4.0
Whiting
Data for 5 surveys over 28 years : 1989-2016
Regression type $=\mathrm{C}$
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as .00
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.
yearclass:2015

| IBTSq11 0.5980 | 11.240 .2580 | 0.743325 | 5.639 | 14.61 | 0.2745 | 0.3622 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| IBTSq12 0.7673 | 10.350 .2067 | 0.818425 | NA | NA | NA | NA |
| IBTSq30 0.6311 | 10.580 .3520 | 0.613023 | 7.646 | 15.40 | 0.3899 | 0.1795 |
| IBTSq31 0.6539 | 10.840 .2285 | 0.788324 | 5.695 | 14.56 | 0.2440 | 0.4583 |
| IBTSq32 0.8437 | 10.420 .2607 | 0.739225 | NA | NA | NA | NA |
| VPA Mean NA | NA NA | A NA 25 | 5 NA | 14.71 | 10.4297 | 0.0000 |


| WAP | $\log$ WAP | int.se |
| :--- | :--- | :--- |
| yearclass:2015 2493212 | 14.73 | 0.1542 |

yearclass:2016
index slope intercept se rsquare n indices prediction se.pred WAP.weights

| IBTSq11 0.5980 | 11.240 .2580 | 0.743325 | NA | NA | NA | NA |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| IBTSq12 0.7673 | 10.350 .2067 | 0.818425 | NA | NA | NA | NA |
| IBTSq30 0.6311 | 10.580 .3520 | 0.613023 | 6.879 | 14.92 | 0.3778 | 1 |
| IBTSq31 0.6539 | 10.840 .2285 | 0.788324 | NA | NA | NA | NA |
| IBTSq32 0.8437 | 10.420 .2607 | 0.739225 | NA | NA | NA | NA |
| VPA Mean NA | NA NA | A NA 25 | 5 NA | 14.71 | 10.4297 | 0 |

$$
\text { WAP } \operatorname{logWAP} \text { int.se }
$$

yearclass:2016 $3005768 \quad 14.92 \quad 0.2837$

### 4.6 North Sea plaice

### 4.6.1 Short term forecast and June advice

At WGNSSK 2016 (ICES 2016), the following short term forecast settings were used:

| Year Class | $\begin{gathered} \text { AGE IN } \\ 2016 \end{gathered}$ | XSA <br> SURVIVORS | RCT3 | $\begin{gathered} \text { GM 1957- } \\ 2013 \end{gathered}$ | ACCEPTED ESTIMATE |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | 2 | 952366 | 704951 | 726298 | XSA survivors |
| 2015 | 1 |  | 907736 | 980962 | RCT3 |
| 2016 | 0 |  |  | 980962 | GM 1957-2013 |

### 4.6.2 New survey information

The new survey information that is available comes from the Beam Trawl Survey (BTS), that was initiated in 1985 and was set up to obtain indices of the younger age groups of plaice and sole, covering the south-eastern part of the North Sea. Since IBPplaice (ICES 2013), the assessment uses the combined BTS-Isis and BTS-Tridens index. This index have a shorter time series due to the BTS-Tridens only starting in 1996.

### 4.6.3 RCT3 Analysis

The RCT3 analysis on the BTS-combined survey indices for ages 1 and 2 was conducted as specified in the Report of the Ad hoc Group on Criteria for Reopening Fisheries Advice (AGCREFA; ICES 2008). Hence, the specifications for the RCT3 were:

| REGRESSION TYPE? | C |
| :--- | :--- |
| Tapered time weighting required? | N |
| Shrink estimates toward mean? | N |
| Exclude surveys with SE's greater than that of mean: | N |
| Enter minimum log S.E. for any survey: | 0.0 |
| Min. no. of years for regression (3 is the default) | 3 |
| Apply prior weights to the surveys? | N |

The input data for the last 42 years including the assessment estimates for the two ages are presented in Table 2.6.1. In 2016, the new data comprises age 1 of year class 2015 and age 2 of year class 2014. The last 4 years from the assessment estimates were removed from the time series.

## Table 2.6.1 North Sea plaice RCT3 input data

| yc | N_Age_1 | N_Age_2 | SNS0 | SNS1 | SNS2 | BTS1 | BTS2 | DFSO |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1974 | 864714 | 548413 | NA | NA | 2402.6 | NA | NA | NA |
| 1975 | 691691 | 448206 | NA | NA | 3423.8 | NA | NA | NA |
| 1976 | 990829 | 649359 | NA | NA | 12678 | NA | NA | NA |
| 1977 | 920713 | 616240 | NA | NA | 9828.8 | NA | NA | NA |
| 1978 | 905430 | 539147 | NA | NA | 12882.3 | NA | NA | NA |
| 1979 | 1148883 | 823274 | NA | NA | 18785.3 | NA | NA | NA |
| 1980 | 901574 | 687932 | NA | NA | 8642 | NA | NA | NA |
| 1981 | 2111275 | 1515904 | NA | NA | 13908.6 | NA | NA | NA |
| 1982 | 1368338 | 988318 | NA | NA | 10412.8 | NA | NA | NA |
| 1983 | 1299663 | 880358 | NA | NA | 13847.8 | NA | NA | NA |
| 1984 | 1880989 | 1316261 | NA | NA | 7580.4 | NA | NA | NA |
| 1985 | 4797263 | 3276305 | NA | NA | 32991.1 | NA | NA | NA |
| 1986 | 1979144 | 1446915 | NA | NA | 14421.1 | NA | NA | NA |
| 1987 | 1830953 | 1325118 | NA | NA | 17810.2 | NA | NA | NA |
| 1988 | 1250820 | 927707 | NA | NA | 7496 | NA | NA | NA |
| 1989 | 1084035 | 841178 | NA | NA | 11247.2 | NA | NA | NA |
| 1990 | 959797 | 692911 | NA | NA | 13841.8 | NA | NA | 439.6 |
| 1991 | 811532 | 599095 | NA | NA | 9685.6 | NA | NA | 332.4 |
| 1992 | 565366 | 416644 | NA | NA | 4976.6 | NA | NA | 180.3 |
| 1993 | 480910 | 374746 | NA | NA | 2796.4 | NA | NA | 217 |
| 1994 | 1197928 | 963465 | NA | NA | 10268.2 | NA | 99.6 | 283.4 |
| 1995 | 1339279 | 1104956 | NA | NA | 4472.7 | 143.9 | 28.7 | 146.1 |
| 1996 | 2212118 | 1878375 | NA | NA | 30242.2 | 386.8 | 177.6 | 619.6 |
| 1997 | 813659 | 636602 | NA | NA | 10272.1 | 131.2 | 53.6 | 229.2 |
| 1998 | 882903 | 677250 | NA | NA | 2493.4 | 117 | 38.9 | NA |


| 1999 | 1035755 | 836291 | NA | 22855 | 2898.5 | 108.4 | 39.8 | NA |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 577074 | 488990 | 24213.5 | 11510.5 | 1102.7 | 80.3 | 26.7 | 124.9 |
| 2001 | 1857519 | 1384282 | 99628 | 30809.2 | NA | 217.3 | 94.4 | 313.2 |
| 2002 | 579018 | 458760 | 31202 | NA | 1349.7 | 53.6 | 30.3 | 122.9 |
| 2003 | 1375294 | 1021614 | NA | 18201.6 | 1818.9 | 101.4 | 45.6 | 238.6 |
| 2004 | 803930 | 635267 | 13537.2 | 10118.4 | 1571 | 70.8 | 42.9 | 126.7 |
| 2005 | 979159 | 675430 | 27390.6 | 12164.2 | 2133.9 | 54.9 | 44.4 | 85.9 |
| 2006 | 1143879 | 959761 | 51124.2 | 14174.5 | 2700.4 | 139.4 | 89.7 | 168 |
| 2007 | 1071558 | 840310 | 40580.9 | 14705.8 | 2018.7 | 98.9 | 76.5 | 98.3 |
| 2008 | 1110840 | 862227 | 50179.3 | 14860 | 1811.5 | 170.8 | 69.5 | 129.7 |
| 2009 | 1419551 | 1123228 | 53258.8 | 11946.9 | 1142.5 | 144.8 | 126 | 141.9 |
| 2010 | 1892434 | 1599807 | 49347.2 | 18348.6 | 2928.6 | 226.5 | 149.6 | 179.6 |
| 2011 | 1274853 | 1065903 | 52643 | 5893.4 | 3021.3 | 118.4 | 90.5 | 93 |
| 2012 | NA | NA | 45027.1 | 15394.9 | 2258.3 | 192.8 | 123.2 | 181.1 |
| 2013 | NA | NA | 44327.5 | 17312.7 | 5040.4 | 155.2 | 156.6 | 168.5 |
| 2014 | NA | NA | 11722.3 | 16726.5 | NA | 116.5 | 68.8 | 108 |
| 2015 | NA | NA | 30494.5 | NA | NA | 112.02 | NA | 100.2 |

### 4.6.4 Update protocol calculations

The outcomes from the RCT3 analyses for the two ages are presented in table 4.6.2.
For age 1 , the D value for this age indicates 0.61 , a small positive signal but not significantly different from spring assumptions. For age 2 the $D$ value $=0.04$ indicating a small positive signal, but not significantly different from spring assumptions. Therefore, a reopening is not warranted.

Table 2.6.2 North Sea plaice RCT3 output for age 1 and 2 and D calculation

## D calculation North Sea plaice age 1

nalysis by RCT3 ver4.0
Plaice

Data for 1 surveys over 42 years : 1974 - 2015
Regression type $=C$
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as . OO
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.
yearclass:2015
index slope intercept se rsquare $n$ indices prediction se.pred
WAP. weights
$\begin{array}{lllllllll}B T S 1 & 0.909 & 9.559 & 0.2699 & 0.6791 & 17 & 4.719 & 13.85 & 0.2959\end{array}$
$\begin{array}{lrrrrrrr}\text { VPA Mean NA } & \text { NA } & \text { NA } & \text { NA } & 38 & \text { NA } & 13.95 & 0.4498\end{array}$
WAP logWAP int.se
yearclass:2015 $1064735 \quad 13.850 .2472$
Spring assumption for age 1: 907736; $\log (907736)=13.72$

| CALCULATIONS FOR 2013 YEAR-CLASS AT AGE 1 |  |  |  |
| :--- | :---: | :---: | :---: |
| Log WAP from RCT3 $(R)$ | 13.85 |  |  |
| Log of recruitment assumed in spring $(A)$ | 13.72 |  |  |
| Int SE of log WAP $(S)$ | 0.25 |  |  |
| $\left(D=\frac{R-A}{S}\right)$ | 0.61 |  |  |
| Distance D |  |  |  |

A positive signal, but not substantially different from spring assumptions.

## D calculation North Sea plaice age 2

```
Analysis by RCT3 ver4.0
Plaice
Data for 1 surveys over 42 years : 1974 - 2015
Regression type = C
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as .00
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.
yearclass:2014
    index slope intercept se rsquare n indices prediction se.pred
WAP.weights
    BTS2 0.8255 10.29 0.301-0.6277 18 4.231 13.78 0.3282
VPA Mean NA NA NA NA 38 NA N NA.65 0.4568 N
    WAP logWAP int.se
yearclass:2014 923678 13.78 0.2666
```

```
Spring assumption for age 2: 952366; log(952366) = 13.77
```

| CALCULATIONS FOR 2012 YEAR-CLASS AT AGE 2 |  |
| :--- | :---: |
| Log WAP from RCT3 $(R)$ | 13.78 |
| Log of recruitment assumed in spring $(A)$ | 13.77 |
| Int SE of log WAP $(S)$ | 0.27 |
| Distance D $\left(D=\frac{R-A}{S}\right)$ | 0.04 |

A negative signal, but not substantially different from spring assumptions.

### 4.6.5 Revised forecast

Since none of the new survey indices indicates a substantial difference in perceived recruitment (compared to the spring assumptions), no further STF was done.

### 4.6.6 Reopening performance

A review of reopened advice since 2009 reveals that the advice was reopened twice: in 2014 and in 2015. In 2014 the recruitment estimate was increased from 936981 to 1309243. In 2015 the recruitment estimate was increased from 650882 to 826318 . For both years, the autumn update of the recruitment corresponded more closely to the assessment outcomes for those ages and years. The absolute difference between recruitment used in the spring forecast and the estimate in 2016 for 2014 and 2015 was $52.3 \%$ and $42.9 \%$, respectively, The difference for the recruitment used in the autumn forecast and the 2016 was smaller, with $33.4 \%$ and $27.5 \%$ respectively.

### 4.6.7 References

ICES. 2008. Report of the Ad hoc Group on Criteria for Reopening Fisheries Advice (AGCREFA). ICES CM 2008/ACOM:60.

ICES. 2013. Report of the Inter-Benchmark Protocol for Plaice in Subarea IV (IBP Plaice), April 2013, By correspondence. ICES CM 2013/ACOM:63. 78 pp.

ICES. 2016. Report of the Working Group for the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), ICES HQ, Copenhagen, Denmark. ICES CM 2016/ACOM.

### 4.7 North Sea sole

### 4.7.1 Short term forecast and June advice

At WGNSSK 2016 (ICES, 2016), the following short term forecast settings were used:

|  |  | AAP |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Year Class | AGe in 2016 | Thousands | RCT3 <br> Thousands | GM(1957-2012) <br> THousANDS |
| 2014 | 2 | 163431 | 137233 | 99023 |
| 2015 | 1 |  | 59248 | 111851 |
| 2016 | Recruit $(0)$ |  |  | 111851 |

### 4.7.2 New survey information

There is new survey information available from the quarter three Beam Trawl Survey (BTS), that was initiated in 1985 and was set up to obtain indices of the younger age groups of plaice and sole.

### 4.7.3 RCT3 Analysis

The RCT3 analysis on the BTS ISIS survey indices for ages 1 and 2 was conducted as specified in the Report of the Ad hoc Group on Criteria for Reopening Fisheries Advice (AGCREFA; ICES 2008). Hence, the specifications for the RCT3 were:

| REGRESSION TYPE? |  |
| :--- | :--- |
| Tapered time weighting required? | C |
| Shrink estimates toward mean? | N |
| Exclude surveys with SE's greater than that of mean: | N |
| Enter minimum log S.E. for any survey: | N |
| Min. no. of years for regression (3 is the default) | 0.0 |
| Apply prior weights to the surveys? | N |

The input data for the last 42 years including the assessment estimates for the two ages are presented in Table 2.6.1. In autumn 2016, the new data derived from the recently conducted surveys comprises age 1 of year class 2015 and age 2 of year class 2014. The last 4 years from the assessment estimates were removed from the time series.

Table 2.7.1. North Sea sole RCT3 input data (shaded cells are new values from 2016 surveys, age 1 and age 2 estimated in separate analysis).
$\left.\begin{array}{llllllll}\text { yc } & \text { N_Age_1 N_Age_2 SNS0 } & \text { SNS1 } & \text { BTS1 } & \text { BTS2 } & \text { DFS0 } \\ 1974 & 59092 & 52591 & 174.4 & 525.4 & \text { NA } & \text { NA } & \text { NA } \\ 1975 & 138917 & 122596 & 577.5 & 1399.4 & \text { NA } & \text { NA } & \text { NA } \\ 1976 & 172911 & 153675 & 464.6 & 3742.9 & \text { NA } & \text { NA } & \text { NA } \\ 1977 & 63381 & 56908 & 1585 & 1547.7 & \text { NA } & \text { NA } & \text { NA } \\ 1978 & 17880 & 16108 & 10370.5 & 93.8 & \text { NA } & \text { NA } & \text { NA } \\ 1979 & 181294 & 163061 & 3922.7 & 4312.9 & \text { NA } & \text { NA } & \text { NA } \\ 1980 & 218640 & 195134 & 5145.8 & 3737.2 & \text { NA } & \text { NA } & \text { NA } \\ 1981 & 204054 & 179807 & 3240.7 & 5856.5 & \text { NA } & \text { NA } & \text { NA } \\ 1982 & 203846 & 179488 & 2147 & 2621.1 & \text { NA } & \text { NA } & \text { NA } \\ 1983 & 95271 & 84715 & 769.1 & 2493.1 & \text { NA } & 7.121 & \text { NA } \\ 1984 & 111974 & 100418 & 3334 & 3619.4 & 7.031 & 5.183 & \text { NA } \\ 1985 & 163489 & 147256 & 2713.4 & 3705.1 & 7.168 & 12.548 & \text { NA } \\ 1986 & 84106 & 75889 & 742 & 1947.9 & 6.973 & 12.512 & \text { NA } \\ 1987 & 686583 & 619725 & 13610.1 & 11226.7 & 83.111 & 68.084 & \text { NA } \\ 1988 & 135009 & 121800 & 522.7 & 2830.7 & 9.015 & 24.487 & \text { NA } \\ 1989 & 247976 & 223426 & 1743.4 & 2856.2 & 37.839 & 28.841 & \text { NA } \\ 1990 & 88638 & 79701 & 50.8 & 1253.6 & 4.035 & 22.284 & 6.38 \\ 1991 & 442592 & 396755 & 3639.7 & 11114 & 81.625 & 42.345 & 167.56 \\ 1992 & 88109 & 78597 & 302.9 & 1290.8 & 6.35 & 7.121 & 9.27 \\ 1993 & 64572 & 57149 & 231.3 & 651.8 & 7.66 & 8.458 & 15.32 \\ 1994 & 113851 & 99840 & 4692.7 & 1362.1 & 28.125 & 7.634 & 22.06 \\ 1995 & 76031 & 66385 & 1374.9 & 218.4 & 3.975 & 4.919 & 7.06 \\ 1996 & 326152 & 285876 & 2322.3 & 10279.3 & 169.343 & 27.422 & 40.27 \\ 1997 & 150896 & 133132 & 803 & 4094.6 & 17.108 & 18.363 & 26.94 \\ 1998 & 116975 & 103693 & 327.9 & 1648.9 & 11.96 & 6.144 & \text { NA } \\ 1999 & 140444 & 124684 & 2187.9 & 1639.2 & 14.594 & 9.963 & \text { NA } \\ 2000 & 72200 & 63920 & 70 & 970.3 & 7.998 & 4.182 & 9.5 \\ 2001 & 217842 & 190844 & 8340 & 7547.5 & 20.989 & 9.947 & 51.42 \\ 2002 & 96079 & 82677 & 1127.7 & \text { NA } & 10.507 & 4.354 & 58.58 \\ 2003 & 52786 & 45003 & \text { NA } & 1369.5 & 4.192 & 3.395 & 10.61 \\ 2014 & \text { NA } & \text { NA } & \text { NA } & 126 & 3192 & 21.099 & 11.982\end{array}\right) 23.62$

### 2.7.4 Update protocol calculations

The outcomes from the RCT3 analyses for the two ages are presented in table 2.6.2.
The D value for age 1 is 0.46 (Table 2.7.2), a weak positive signal, not significantly different from the spring assumption $(\mathrm{D}<1)$. For age 2 the D value is -0.74 (Table 2.7.2), a weak negative signal, not significantly different from the spring assumption $(\mathrm{D}<1)$. Hence, the short term forecast does not need to be re-run.

Table 2.7.2 North Sea sole RCT3 output for age 1 and 2 and D calculation

## D calculation North Sea sole age 1

Analysis by RCT3 ver4.0: Sole
Data for 1 surveys over 42 years : 1974-2015
Regression type $=\mathrm{C}$
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as . 00
Minimum of 3 points used for regression

Forecast/Hindcast variance correction used.
yearclass:2015
index slope intercept se rsquare n indices prediction se.pred WAP.weights $\begin{array}{llllllll}B T S 1 & 0.759 & 9.754 & 0.3838 & 0.7476 & 28 & 1.842 & 11.15 \\ 0.4099 & 1\end{array}$

VPA Mean NA NA NA NA 38 NA $11.710 .6827 \quad 0$

WAP $\operatorname{logWAP~int.se~}$
yearclass:2015 6969111.150 .3514

Spring assumption for age $1: 59248 ; \log (59248)=10.99$

| CALCULATIONS FOR 2013 YEAR-CLASS AT AGE 1 |  |
| :--- | :---: |
| Log WAP from RCT3 $(R)$ | 11.15 |
| Log of recruitment assumed in spring $(A)$ | 10.99 |
| Int SE of log WAP $(S)$ | 0.35 |
| Distance D $\left(D=\frac{R-A}{S}\right)$ | 0.46 |

## D calculation North Sea sole age 2

Analysis by RCT3 ver4.0: Sole

Data for 1 surveys over 42 years : 1974-2015
Regression type $=C$
Tapered time weighting not applied
Survey weighting not applied
Final estimates not shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as .00
Minimum of 3 points used for regression

Forecast/Hindcast variance correction used.
yearclass:2014
index slope intercept se rsquare $n$ indices prediction se.pred WAP.weights
$\begin{array}{llllllll}\text { BTS2 } & 0.9157 & 9.463 & 0.3869 & 0.7349 & 30 & 2.483 & 11.74 \\ 0.4073 & 1\end{array}$
VPA Mean NA NA NA NA 39 NA 11.59 0.6760 0

## WAP $\operatorname{logWAP~int.se~}$

yearclass:2014 12505711.740 .3489

Spring assumption for age 2: $163431 ; \log (163431)=12.00$

| CALCuLATIONS FOR 2013 Year-CLASS AT AGE 2 |  |
| :--- | :--- |
| Log WAP from RCT3 $(R)$ | 11.74 |
| Log of recruitment assumed in spring $(A)$ | 12.00 |
| Int SE of log WAP $(S)$ | 0.35 |
| Distance D $\left(D=\frac{R-A}{S}\right)$ | -0.74 |

### 2.7.5 Reopening performance

A review of reopened advice since 2009 reveals that the advice was reopened twice: in 2014 and in 2015. In 2014 the age 2 estimate was increased from 54268 to 65474. In 2015 the recruitment estimate was increased from 103741 to 135220 . For both years, the autumn update of the recruitment corresponded more closely to the assessment outcomes for those ages and years. The absolute difference between recruitment used in the
spring forecast and the estimate in 2016 for 2014 and 2015 was 53.5 and $46.6 \%$, respectively, The difference for the recruitment used in the autumn forecast and the 2016 was smaller, with $43.9 \%$ and $30.3 \%$ respectively.

### 2.7.6 References

ICES. 2008. Report of the Ad hoc Group on Criteria for Reopening Fisheries Advice (AGCREFA). ICES CM 2008/ACOM:60.

ICES. 2016. Report of the Working Group for the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), 2016, ICES HQ, Copenhagen, Denmark. ICES CM 2016/ACOM.

### 4.8 North Sea Nephrops

## Nephrops FU6

The annual underwater TV survey of the Farn Deeps area was undertaken $21^{\text {st }}-28$ th June 2016.

The survey was completed without any technical issues and the visibility was excellent. All 110 stations were completed with valid counts generated using the standard protocols for counting and quality assurance. The survey results show the usual pattern to the densities as with previous years of a higher density spine down the western side of the ground. There does appear to be a slight southward shift in stock densities with an absence of Nephrops in the far north and a higher relative abundance in the far south (figure 1).

Total abundance in 2016 is estimated to be 697 million with a $95 \% \mathrm{CI}$ of 19 million. The advice in June 2016 was based upon the 2015 survey which showed 565 million with a $95 \%$ CI of 13 million. The increase in abundance from 2015 to 2016 was 132 million, well beyond the confidence envelope of the 2015 survey (table 1 and figure 2).

## It is therefore recommended that the advice be reopened.

Btrigger for FU6 Nephrops is 858 million and the MSY harvest rate is $8.1 \%$. Following ICES procedures, when the abundance index is below Btrigger, the target harvest rate is reduced linearly by the ratio of current abundance: Btrigger. This results in a target Harvest Rate of $6.6 \%$.

Mean weights and discard rates have not been updated (as this update only has new survey data), so the updated advice is only based upon the change in abundance estimate.

Nephrops in IV: FU6 updated survey results. Table 1 survey result history

| Year | Stations | Season | Mean density | Absolute <br> Abundance | 95\% CONFIDENCE INTERVAL | Method |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | burrows/m ${ }^{2}$ | millions | millions |  |
| 1997 | 87 | Autumn | 0.46 | 1500 | 125 | Box |
| 1998 | 91 | Autumn | 0.33 | 1090 | 89 | Box |
| 1999 | - | Autumn | No survey |  |  | Box |
| 2000 | - | Autumn | No survey |  |  | Box |
| 2001 | 180 | Autumn | 0.56 | 1685 | 67 | Box |
| 2002 | 37 | Autumn | 0.33 | 1048 | 112 | Box |
| 2003 | 73 | Autumn | 0.33 | 1085 | 90 | Box |
| 2004 | 76 | Autumn | 0.43 | 1377 | 101 | Box |
| 2005 | 105 | Autumn | 0.49 | 1657 | 148 | Box |
| 2006 | 105 | Autumn* | 0.37 | 1244 | 114 | Box |
| 2007 | 105 | Autumn* | 0.28 | 858 | 23 | Geostatistics |
| 2008 | 95 | Autumn* | 0.31 | 987 | 39 | Geostatistics |
| 2009 | 76 | Autumn* | 0.22 | 682 | 38 | Geostatistics |
| 2010 | 95 | Autumn* | 0.25 | 785 | 21 | Geostatistics |
| 2011 | 97 | Autumn* | 0.28 | 878 | 17 | Geostatistics |
| 2012 | 97 | Autumn* | 0.24 | 758 | 13 | Geostatistics |
| 2013 | 110 | Summer | 0.23 | 706 | 18 | Geostatistics |
| 2014 | 110 | Summer | 0.24 | 755 | 18 | Geostatistics |
| 2015 | 110 | Summer | 0.18 | 565 | 13 | Geostatistics |
| 2016 | 110 | Summer | 0.22 | 697 | 19 | Geostatistics |

## Nephrops in IV: FU6 updated survey results. Table 2 revised catch advice tables

Catch options assuming zero discards

| RATIONALE | BASIS | Total CATCH | WANTED <br> CATCH* | UNWANTED <br> CATCH* | HARVEST <br> RATE** |
| :--- | :--- | :--- | :--- | :--- | :--- |
| MSY <br> approach | MSY approach | 1125 | 1004 | 121 | $6.60 \%$ |
| Other <br> options | FMSY | Fcurrent (2013-2015) | 2641 | 1385 | 2357 |

* Wanted" and "unwanted" catch are used to described Norway lobster that would be landed and discarded in the absence of the EU landing obligation based on discard rates estimates for average (20132015).
** Calculated for dead removals and applied to total catch.

Discarding assumed below MCS only*

|  |  | TOTAL CATCH | Dead REMOVALS | LANDINGS (Wanted CATCH) | UnWANTED >MCS** | Dead <br> DISCARDS $<\mathrm{MCS}$ | Surviving DISCARDS | Harvest RATE*** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Rationale | BASIS | L+U+DD+SD | L+U+DD | L | U | DD | SD | for $\mathrm{L}+\mathrm{U}+\mathrm{DD}$ |
| MSY <br> approach | MSY approach | 1143 | 1138 | 1020 | 90 | 28 | 5 | 6.60\% |
| Other options | FMSY | 1407 | 1401 | 1256 | 111 | 35 | 6 | 8.12\% |
|  | Fcurrent (2013-2015) | 2683 | 2671 | 2394 | 211 | 66 | 12 | 15.48\% |

* Assumed for all fleets
** Unwanted landings (U) are those animals $>$ MCS but historically discarded
*** Calculated for dead removals


Nephrops in IV: FU6 survey update. Figure 1. Abundance maps.


Nephrops in IV: FU6 survey update. Figure 2. Abundance time series.

## Fladen (FU7)

The most recent UWTV survey for this stock was carried out in June 2016. The survey followed the usual procedures for Scottish UWTV surveys, and these are described in more detail in the Stock Annex.

The UWTV estimate of abundance used in the June 2016 advice and based on the 2015 survey is 2569 million with a $95 \%$ CI of 320 million (Table 1; Figure $1 \& 2$ ). The estimate from the 2016 summer survey is 4449 million ( $73 \%$ increase on the 2015 value). The 2016 value is significantly different from that of 2015 (ACOM specifies 1 SD, this is well over the specified threshold) and therefore the advice for FU7 may be reopened.

The advice for 2017 for Category 1 stocks (where assessment includes landings and discards data) is based on catches. The catch prediction for 2017 under the landing obligation and assuming discarding below MCS only (Table 2) following the MSY approach is 11852 tonnes (the June advice was 6844 tonnes). Mean weights and discard rates have not been revised in October 2016 (as this update only has new 2016 summer survey data), so the update of the advice is only due to the change in the abundance estimate. Discards survival for Nephrops in FU7 is assumed to be $25 \%$. ICES was also requested to provide a second catch options table assuming zero discards (Table 3).

The large abundance increase in 2016 may be be related with a strong recruitment event. The size of Nephrops burrows is not quantified in the TV surveys but burrow counters participating in the last survey reported a large number of small burrows in FU 7. If the increase is dominated by small Nephrops, it is unlikely they will appear in the fishery straight away given the selectivity observed for this FU. It should be noted that in recent years the catch in this FU has been lower than advised. The increase in the 2016 abundance is substantial and this translates into a large increase in the advice for 2017 compared with that released in June. In the event that this advice is updated it should be emphasized that if the large difference between advice and catches remains, it may be transferred to other FUs in the North Sea which could result in nonprecautionary exploitation of those FUs.

Table 1. Nephrops, Fladen (FU 7): Results of the 1992-2016 TV surveys.

| Year | Stations | Abundance | Mean density | 95\% CONFIDENCE INTERVAL |
| :---: | :---: | :---: | :---: | :---: |
|  |  | millions | burrows/m2 | millions |
| 1992 | 69 | 3661 | 0.13 | 376 |
| 1993 | 74 | 4450 | 0.16 | 569 |
| 1994 | 59 | 6170 | 0.22 | 814 |
| 1995 | 61 | 4987 | 0.18 | 896 |
| 1996 | No survey |  |  |  |
| 1997 | 56 | 2767 | 0.10 | 510 |
| 1998 | 60 | 3838 | 0.13 | 717 |
| 1999 | 62 | 4146 | 0.15 | 649 |
| 2000 | 68 | 3628 | 0.13 | 491 |
| 2001 | 50 | 4981 | 0.17 | 970 |
| 2002 | 54 | 6087 | 0.21 | 757 |
| 2003 | 55 | 5547 | 0.20 | 1076 |
| 2004 | 52 | 5725 | 0.20 | 1030 |
| 2005 | 72 | 4325 | 0.16 | 662 |
| 2006 | 69 | 4862 | 0.17 | 619 |
| 2007 | 82 | 7017 | 0.25 | 730 |
| 2008 | 74 | 7360 | 0.26 | 1019 |
| 2009 | 59 | 5457 | 0.19 | 772 |
| 2010 | 67 | 5224 | 0.19 | 710 |
| 2011 | 73 | 3382 | 0.12 | 435 |
| 2012 | 70 | 2748 | 0.10 | 392 |
| 2013 | 71 | 2902 | 0.10 | 335 |
| 2014 | 70 | 2990 | 0.11 | 412 |
| 2015 | 71 | 2569 | 0.091 | 320 |
| 2016 | 78 | 4449 | 0.158 | 662 |

## FU7 basis for the catch options

| VARIABLE | VALUE | SourCE | Notes |
| :--- | :--- | :--- | :--- |
| Stock abundance | 4449 <br> million <br> individuals | ICES <br> (2016a) | UWTV 2015 |
| Mean weight in <br> landings | 38.24 g | ICES <br> (2016a) | Average 2013-2015 |
| Mean weight in <br> discards | 15.30 g | ICES <br> (2016a) | Average 2013-2015 |
| Mean weight in <br> unwanted catch >MCS | 16.13 g | ICES <br> (2016a) | Average 2013-2015 |
| Mean weight in <br> unwanted catch <MCS | 7.58 g | ICES <br> (2016a) | Average 2013-2015 |
| Discard rate (total) | $0.83 \%$ | ICES <br> (2016a) | Average 2013-2015 (proportion by <br> number) |
| Discard rate (>MCS) | $0.75 \%$ | ICES <br> (2016a) | Average 2013-2015 (proportion by <br> number) |
| Discard rate (<MCS) | $0.08 \%$ | ICES <br> (2016a) | Average 2013-2015 (proportion by <br> number) |
| Discard survival rate | $25 \%$ | ICES <br> (2016a) | Proportion by number, only applies in <br> scenarios when discarding is allowed. |
| Dead discard rate <br> (total) | $0.62 \%$ | ICES <br> (2016a) | Average 2013-2015 (proportion by <br> number), only applies in scenarios when <br> discarding is allowed. |
| Dead discard rate <br> (<MCS) | $0.12 \%$ | ICES |  |
| (2016a) | Average (proportion by number) 2013- <br> 2015, only applies in scenarios when <br> discarding is allowed below MCS. |  |  |

Table 2. Revised Advice table assuming discarding below MCS only*

|  |  | Total catch | Dead removals | Landings (Wanted CATCH) | UnWANTED $>\text { MCS** }^{*}$ | Dead DISCARDS <MCS | Surviving DISCARDS | Harvest RATE*** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Rationale | Basis | L+U+DD+SD | L+U+DD | L | U | DD | SD | for L+U+DD |
| MSY <br> approach | MSY approach | 11852 | 11852 | 11813 | 38 | 1 | 0 | 7.0\% |
| Other options | FMSY | 12699 | 12698 | 12656 | 40 | 2 | 1 | 7.5\% |
|  | F2015 | 3386 | 3386 | 3375 | 11 | 0 | 0 | 2.0\% |
|  | F2013-2015 | 4911 | 4911 | 4894 | 16 | 1 | 0 | 2.9\% |
|  | F35\%SpR | 18963 | 18962 | 18900 | 60 | 2 | 1 | 11.2\% |
|  | Fmax | 27767 | 27766 | 27675 | 88 | 3 | 1 | 16.4\% |

[^10]Table 3. Revised Advice table assuming zero discards

| Rationale | BASIS | Total CATCHES | Wanted CATCHES* | UNWANTED CATCHES* | Harvest RATE** |
| :---: | :---: | :---: | :---: | :---: | :---: |
| MSY <br> approach | MSY approach | 11850 | 11810 | 40 | 7.0\% |
| Other options | Fmsy | 12696 | 12654 | 42 | 7.5\% |
|  | F2015 | 3385 | 3374 | 11 | 2.0\% |
|  | F2013-2015 | 4909 | 4893 | 16 | 2.9\% |
|  | F35\%SpR | 18959 | 18896 | 63 | 11.2\% |
|  | Fmax | 27763 | 27670 | 93 | 16.4\% |

* "Wanted" and "unwanted" catch are used to described Nephrops that would be landed and discarded in the absence of the EU landing obligation based on discard rates estimates for average (2013-2015).
** Calculated for dead removals and applied to total catch.


Figure 1. Nephrops, Fladen (FU 7). TV survey distribution and relative density in 2016. Green and brown areas represent areas of suitable sediment for Nephrops. Density proportional to circle radius. Red crosses represent zero observations.


Figure 2. Nephrops, Fladen (FU 7): Results of the 1992-2016 TV surveys.

## Firth of Forth (FU8)

The most recent UWTV survey for this stock was carried out in August 2016. The survey followed the usual procedures for Scottish UWTV surveys, and these are described in more detail in the Stock Annex.

The UWTV estimate of abundance used in the June 2016 advice and based on the 2015 survey is 664 million with a $95 \% \mathrm{CI}$ of 127 million (Table 4; Figure $3 \& 4$ ). The estimate from the 2016 summer survey is 797 million ( $20 \%$ increase on the 2015 value). The 2015 value is significantly different from that of 2015 (ACOM specifies 1 SD, this is over 2 SD) and therefore the advice for FU8 may be reopened.

The advice for 2017 for Category 1 stocks (where assessment includes landings and discards data) is based on catches. The catch prediction for 2017 under the landing obligation and assuming discarding below MCS only (Table 5) following the MSY approach is 2548 tonnes (the June advice was 2123 tonnes). Mean weights and discard rates have not been revised in October 2016 (as this update only has new 2016 summer survey data), so the update of the advice is only due to the change in the abundance estimate. Discards survival for Nephrops in FU8 is assumed to be $25 \%$. ICES was also requested to provide a second catch options table assuming zero discards (Table 6).

Table 4. Nephrops, Firth of Forth (FU 8): Results of the 1993-2016 TV surveys.

| Year | Stations | Mean Density | Abundance | 95\% CONF INTERVAL |
| :---: | :---: | :---: | :---: | :---: |
|  |  | burrows/m ${ }^{2}$ | millions | millions |
| 1993 | 37 | 0.61 | 555 | 142 |
| 1994 | 30 | 0.49 | 448 | 78 |
| 1995 | no survey |  |  |  |
| 1996 | 27 | 0.41 | 375 | 88 |
| 1997 | no survey |  |  |  |
| 1998 | 32 | 0.32 | 292 | 81 |
| 1999 | 49 | 0.51 | 463 | 78 |
| 2000 | 53 | 0.48 | 443 | 70 |
| 2001 | 46 | 0.46 | 419 | 79 |
| 2002 | 41 | 0.56 | 508 | 119 |
| 2003 | 36 | 0.84 | 767 | 138 |
| 2004 | 37 | 0.69 | 630 | 141 |
| 2005 | 54 | 0.78 | 710 | 143 |
| 2006 | 43 | 0.91 | 827 | 125 |
| 2007 | 49 | 0.76 | 692 | 132 |
| 2008 | 38 | 0.97 | 881 | 297 |
| 2009 | 45 | 0.80 | 732 | 142 |
| 2010 | 39 | 0.75 | 682 | 147 |
| 2011 | 45 | 0.58 | 533 | 87 |
| 2012 | 66 | 0.57 | 522 | 64 |
| 2013 | 51 | 0.73 | 668 | 125 |
| 2014 | 51 | 0.47 | 428 | 80 |
| 2015 | 51 | 0.73 | 664 | 127 |
| 2016 | 50 | 0.87 | 797 | 146 |

FU8 basis for the catch options

| VARIABLE | VALUE | SourCe | Notes |
| :--- | :--- | :--- | :--- |
| Stock abundance | 797 million <br> individuals | ICES <br> $(2016 a)$ | UWTV 2015 |
| Mean weight in landings | 21.81 g | ICES <br> $(2016 a)$ | Average 2013-2015 |
| Mean weight in discards | 10.74 g | ICES <br> $(2016 a)$ | Average 2013-2015 |
| Mean weight in unwanted <br> catch $>$ MCS | 13.71 g | ICES <br> $(2016 a)$ | Average 2013-2015 |
| Mean weight in unwanted <br> catch <MCS | 7.25 g | ICES <br> $(2016 a)$ | Average 2013-2015 |
| Discard rate (total) | $24.9 \%$ | ICES <br> $(2016 a)$ | Average 2013-2015 (proportion by <br> number) |
| Discard rate (>MCS) | $13.5 \%$ | ICES <br> $(2016 a)$ | Average 2013-2015 (proportion by <br> number) |


| Discard rate (<MCS) | $11.4 \%$ | ICES <br> $(2016 a)$ | Average 2013-2015 (proportion by <br> number) |
| :--- | :--- | :--- | :--- |
| Discard survival rate | $25 \%$ | ICES <br> $(2016 \mathrm{a})$ | Average 2013-2015 (proportion by <br> number), only applies in scenarios <br> when discarding is allowed. |
| Dead discard rate (<MCS) | $8.8 \%$ | ICES <br> $(2016 \mathrm{a})$ | Average (proportion by number) <br> 2013-2015, only applies in scenarios <br> when discarding is allowed below <br> MCS. |

Table 5. Revised Advice table assuming discarding below MCS only*

| Rationale | BASIS | Total catchL+U+DD+SD | $\begin{gathered} \text { DeAD } \\ \text { REMOVALS } \end{gathered}$ | Landings (wanted CATCH) <br> L | UnWANTED $>\text { MCS** }^{* *}$ <br> U | Dead DISCARDS <MCS DD | Surviving DISCARDS SD | $\begin{aligned} & \text { HARVEST } \\ & \text { RATE*** } \\ & \text { for } \\ & \text { L+U+DD } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
| MSY <br> approach | MSY approach | 2548 | 2520 | 2190 | 247 | 83 | 28 | 16.3\% |
| Other options | F0.1 | 1470 | 1454 | 1263 | 143 | 48 | 16 | 9.4\% |
|  | F35SpR | 1987 | 1965 | 1707 | 193 | 65 | 22 | 12.7\% |
|  | F2015 | 2625 | 2597 | 2257 | 255 | 85 | 28 | 16.8\% |
|  | F2013-2015 | 3205 | 3170 | 2755 | 311 | 104 | 35 | 20.5\% |

Table 6. Revised Advice table assuming zero discards

| RATIONALE | BASIS | TOTAL <br> CATCHES | WANTED <br> CATCHES* | UNWANTED <br> CATCHES* | HARVEST <br> RATE** |
| :--- | :--- | :--- | :--- | :--- | :--- |
| MSY approach | MSY approach | 2475 | 2128 | 347 | $16.3 \%$ |
| Other options | F0.1 | 1427 | 1227 | 200 | $9.4 \%$ |
|  | F35SpR | 1929 | 1658 | 271 | $12.7 \%$ |
|  | F2015 | 2551 | 2193 | 358 | $16.8 \%$ |
|  | F2013-2015 | 3113 | 2676 | 437 | $20.5 \%$ |

## 2016



Figure 3. Nephrops, Firth of Forth (FU 8). TV survey distribution and relative density in 2016. Green and brown areas represent areas of suitable sediment for Nephrops. Density proportional to circle radius. Red crosses represent zero observations.
firth forth


Figure 4. Nephrops, Firth of Forth (FU 8): Results of the 1992-2016 TV surveys.

## Moray Firth (FU9)

The most recent UWTV survey for this stock was carried out in August 2016. The survey followed the usual procedures for Scottish UWTV surveys, and these are described in more detail in the Stock Annex.

The UWTV estimate of abundance used in the June 2016 advice and based on the 2015 survey is 347 million with a $95 \%$ CI of 84 million (Table 7; Figure 5 \& 6). The estimate from the 2016 summer survey is 388 million ( $12 \%$ increase on the 2015 value). The 2016 value is just within 1 SD of the 2015 abundance estimate and therefore the advice for FU9 should not be reopened.

Table 7. Nephrops, Moray Firth (FU 9): Results of the 1993-2015 TV surveys.

|  |  | Mean |  | 95\% |
| :---: | :---: | :---: | :---: | :---: |
|  |  | density |  | confidence |
|  |  |  | Abundance | interval |
| Year | Stations | burrows/m² | millions | millions |
| 1993 | 31 | 0.16 | 345 | 78 |
| 1994 | 29 | 0.32 | 702 | 176 |
| 1995 | no survey |  |  |  |
| 1996 | 27 | 0.21 | 465 | 90 |
| 1997 | 34 | 0.12 | 262 | 55 |
| 1998 | 31 | 0.15 | 323 | 95 |
| 1999 | 52 | 0.18 | 400 | 87 |
| 2000 | 44 | 0.17 | 386 | 98 |
| 2001 | 45 | 0.16 | 345 | 112 |
| 2002 | 31 | 0.24 | 521 | 121 |
| 2003 | 32 | 0.33 | 730 | 314 |
| 2004 | 42 | 0.29 | 626 | 186 |
| 2005 | 42 | 0.40 | 869 | 198 |
| 2006 | 50 | 0.21 | 445 | 124 |
| 2007 | 40 | 0.24 | 531 | 156 |
| 2008 | 45 | 0.21 | 481 | 151 |
| 2009 | 50 | 0.19 | 415 | 140 |
| 2010 | 43 | 0.18 | 406 | 116 |
| 2011 | 37 | 0.17 | 372 | 160 |
| 2012 | 44 | 0.14 | 299 | 90 |
| 2013 | 55 | 0.21 | 469 | 106 |
| 2014 | 52 | 0.15 | 331 | 90 |
| 2015 | 52 | 0.16 | 347 | 84 |
| 2016 | 53 | 0.18 | 388 | 87 |

2016


Figure 5. Nephrops, Moray Firth (FU 9). TV survey distribution and relative density in 2016. Green and brown areas represent areas of suitable sediment for Nephrops. Density proportional to circle radius. Red crosses represent zero observations.
moray firth


Figure 6. Nephrops, Moray Firth (FU 9): Results of the 1992-2016 TV surveys.

## Annex 05 List of Stock Annexes

The table below provides an overview of the WGBFAS Stock Annexes. Stock Annexes for other stocks are available on the ICES website Library under the Publication Type "Stock Annexes". Use the search facility to find a particular Stock Annex, refining your search in the left-hand column to include the year, ecoregion, species, and acronym of the relevant ICES expert group.

| Stock ID | Stock Name | Modified | Link |
| :---: | :---: | :---: | :---: |
| bll-nsea_SA | Brill in Subarea 4 and divisions 3.a and 7.d-e | 07-12-2015 13:46 | bll- |
|  |  |  | nsea SA.docx |
| cod-347d_SA | Cod in Subarea 4 and divisions 7.d and 3.a West | 29-06-2015 16:27 | cod- |
|  |  |  | 347d SA.doc |
| dab-nsea_SA | Dab in Subarea 4 and Division 3.a | 04-07-2016 13:59 | dab- |
|  |  |  | nsea SA.docx |
| fle-nsea_SA | Flounder in Subarea 4 and Division 3.a | 29-06-2015 16:29 | $\underline{ }$ |
|  |  |  | nsea SA.docx |
| gug-347d_SA | Grey gurnard in Subarea 4 and divisions 7.d and 3.a | 28-10-2014 13:28 | gug- |
|  |  |  | 347d SA.docx |
| had-346a_SA | Haddock in Subarea 4 and divisions 6.a and 3.a West | 28-10-2014 13:26 | had- |
|  |  |  | 346a_SA.docx |
| lem-nsea_SA | Lemon sole in Subarea 4 and divisions 3.a and 7.d | 29-06-2015 16:30 | lem- |
|  |  |  | nsea SA.docx |
| mur-347d_SA | Striped red mullet in Subarea 4 and divisions 7.d and 3.a | 14-04-2015 14:57 |  |
|  |  |  | 347d SA.docx |
| nep-10_SA | Norway lobster in Division 4.a, FU 10 | 04-11-2014 12:04 | nep- |
|  |  |  | $10 \text { SA.docx }$ |
| nep-32_SA | Norway lobster in Division 4.a, FU 32 | 04-11-2014 12:03 | nep- |
|  |  |  | 32.SA.docx |
| nep-33_SA | Norway lobster in Division 4.b, FU 33 | 05-07-2016 12:11 | nep- |
|  |  |  | $33 \text { SA.docx }$ |
| nep-34_SA | Norway lobster in Division 4.b, FU 34 | 04-11-2014 12:03 |  |
|  |  |  | 34_SA.docx |
| nep-3-4_SA | Norway lobster in Division 3.a | 29-06-2015 16:32 | nep-3- |
|  |  |  | $4 \text { SA.docx }$ |
| nep-5_SA | Norway lobster in divisions 4.b and 4.c, FU 5 | 06-07-2016 09:23 |  |
|  |  |  | $5 \text { SA.docx }$ |
| nep-6_SA | Norway lobster in Division 4.b, FU 6 | 04-11-2014 12:02 |  |
|  |  |  | $6 \text { SA. docx }$ |
| nep-7_SA | Norway lobster in Division 4.a, FU 7 | 29-06-2015 16:38 | nep- |
|  |  |  | 7 SA.docx |
| nep-8_SA | Norway lobster in Division 4.b, FU 8 | 04-11-2014 12:02 |  |
|  |  |  | 8_SA.docx |
| nep-9_SA | Norway lobster in Division 4.b, FU 9 | 04-11-2014 12:01 |  |
|  |  |  | $9 \text { SA.docx }$ |
| nop-34_SA | Norway pout in Subarea 4 and Division 3.a | 04-11-2014 14:26 | nop- |
|  |  |  | $34 \text { SA.docx }$ |
| ple-eche_SA | Plaice in Division 7.d | 29-06-2015 16:40 | ple- |
|  |  |  | eche SA.docx |
| ple-kask_SA |  | 04-11-2014 15:03 |  |
|  |  |  | kask_SA.docx |


| ple-nsea_SA | Plaice in Subarea 4 (North Sea) and Division 3.a (Skagerrak) | 04-11-2014 15:02 | plensea SA.docx |
| :---: | :---: | :---: | :---: |
| pol-nsea_SA | Pollack in Subarea 4 and Division 3.a | 04-11-2014 15:02 | pol- <br> nsea_SA.docx |
| sai-3a46_SA | Saithe in subareas 4-5 and Division 3.a | 06-07-2016 15:52 | sai- <br> 3 346 SA.docx |
| sol-eche_SA | Sole in Division 7.d | 04-11-2014 12:47 | soleche SA.docx |
| sol-nsea_SA | Sole in Subarea 4 | 14-04-2015 14:58 | sol- <br> nsea_SA.docx |

## Annex 6 Benchmark Planning and Data Problems by Stock

Part A
Benchmarks planning WGNSSK

## A.1.1 Latest benchmark results

The ICES Benchmark Workshop on North Sea Stocks (WKNSEA-2016) convened at two meetings in Copenhagen, one data compilation workshop (23-25 November 2015) and the final benchmark meeting (14-18 March 2016).

In WKNSEA-2016, two stocks were benchmarked: Saithe in IV, VI and IIIa, and Dab in IV and IIIa. Furthermore, an inter-benchmark meeting was held by correspondence just prior to WGNSSK-2016 for Whiting in IV and VIId. The most important conclusions for each stock are given below.

Saithe in IV, VI and IIIa (WKNSEA-2016):
A benchmark for North Sea saithe (Subareas 4 and 6, Division 3a) took place in March 2016. The official data call requested landings and discards to be uploaded by métier, quarter, and area into InterCatch; tagging or data that could assist in stock differentiation; maturity and national survey indices not held within DATRAS; national cpue indices; data on natural mortality; and recreational fishery catches by age, year, and quarter.

## Landings and discards from commercial fleets

Discards were raised for all fleets that did not submit information. Discards were raised two ways: treating Norwegian fleets as operating like other fleets and treating Norwegian OTB_DEF fleets as operating like French and German OTB_DEF fleets. A bug in InterCatch resulted in the re-raising of data after the WG met in May for four years. To estimate discards weights- and numbers-at-age prior to 2002, a constant ratio landings/discards by age was applied. Discard weights for age $8+$ were set to 1 .

Ages were allocated using 3 different stratification scenarios: no stratification, by area, and by quarter and area. The 'by quarter and area' stratification was used for the final data set. Discard age allocations were unstratified for most years due to lack of samples; it was only the most recent years which had enough samples to allow for the 'by area and quarter' stratification.

## Survey data

The North Sea IBTS Q1 and Q3 and Scottish West Coast Q1 and Q4 surveys were investigated to see if they could be used as a survey index. A delta-GAM approach was initially used to estimate the index for ages 1-10+, but not all ages were included in the assessment model. At the benchmark, the North Sea IBTS Q1 (ages 3-5) and North Sea IBTS Q3 (ages 3-8), estimated using the delta-GAM approach (Berg et al. 2014), were used in the assessment. It was after an external review post-WGNSSK (end of May) that only the NS-IBTS Q3 (ages 3-8), estimated as the standard DATRAS output, was used. The Q1 survey was eliminated due to movement of fish in and out of the survey area, unrelated to abundance. The delta-GAM approach for estimating the indices was also removed because the method uses one age-length key for all years, whereas the survey data were showing year effects existed.

## Changes to other input data

The benchmark explored an alternate rate for natural mortality rate of 0.26 based on longevity (Then et al. 2014), but opted to remain with the 0.2 value (noting that this deserved exploration in the near future).

Stock weights, estimated from the NS-IBTS Q1 and Q3 and the Scottish West Coast Q1 and Q4 surveys, were used initially in the benchmark model. These were replaced with catch weights for all ages following the external review post-WGNSSK.

The maturity ogive was revised using all available survey data (NS-IBTS and SWCIBTS) from the 1 st and 4th quarters. After much discussion, it was agreed that the ogive including cohort showed too much variability that was unlikely over such a short time period, even after smoothing was applied within the SAM model. Therefore the newly estimated static ogive was used, with some modification based on expert knowledge within the group. This modification was to use a slightly conservative approach for setting the proportions mature at age 3 and 4 to less than that estimated by the model. This was because of variability in the amount of age 3 and 4 year old fish that migrate into the survey area varies annually and those that are present tend to be the larger (and thus faster maturing) fish.

## Combined cpue index

Data for the Norwegian, French, and German commercial trawler fleets were combined into one standardized CPUE index, which is then tuned to the exploitable biomass. A single combined index was estimated to avoid using the same information twice: the information in the catch-at-age matrix and in the three individual cpue fleets in the assessment. There were concerns that using the information twice gave too much weight in the tuning.

## Assessment mode/

A state-space assessment model (SAM; Nielsen and Berg 2014) was used, where the correlation between age groups within years was included in the model, following the method of Berg and Nielsen (2016). The final model includes the standard DATRAS IBTS Q3 index (ages 3-8), a combined cpue index tuned to the exploitable biomass, a new maturity ogive, discard information, and revised catch data for 2002-2015.

## References

Berg, CW, Nielsen, A. 2016. Accounting for correlated observations in an age-based state-space stock assessment model. ICES Journal of Marine Science. doi:10.1093/icesjms/fsw046

Berg, C., Nielsen, A., Christensen, K., 2014. Evaluation of alternative age-based methods for estimating relative abundance from survey data in relation to assessment models. Fisheries Research 151: 91-99.

Nielsen, A. and C.W. Berg. 2014. Estimation of time-varying selectivity in stock assessments using state-space models. Fisheries Research, 158: 96-101.

Then, A., Hoenig, J. M., Hall, N.,G., Hewitt, D.,A., (2014). Evaluating the predictive performance of empirical estimators of natural mortality rate using information on over 200 fish species. ICES Journal of Marine Science 72 (1) 82-92.

## Dab in IV and IIIa (WKNSEA-2016)

A benchmark assessment for North Sea dab was carried out in spring 2016 (14.03. 18.03.2016, ICES HQ Copenhagen). The official data call for this benchmark requested landings and discard data from 2002 - 2014. All relevant countries loaded up these data into the InterCatch data portal which was subsequently used to raise discards and to estimate the total catch for dab. Besides landing and discard data also biological sampling data such as age readings and weight at age data were uploaded and were analysed during the benchmark. The suitability of these data to set up an age based assessment for dab were tested. Further, the suitability of different survey data and indices was tested to set up a survey based assessments model (SURBA).

## Data on landings, discards and biological sampling from commercial fleets

The amount of imported discard data was lower for the earlier years of the time series and the InterCatch raising procedure introduces some uncertainty. However, in general discard data for the most important fleet, the Dutch TBB_DEF_70-99_0_0_all metier, was provided and most of the amount of discard estimates was therefore based on imported data. Thus, it was possible to create a time series on total catch for the years 2002 to 2014. The discard ratio increased from $76 \%$ in 2002 up to $91 \%$ for the last year of this time series. Some countries provided also catch at age (numbers) and weight at age data based on samples from commercial fleets. The age distribution pattern seems to be odd for the years 2002 - 2005 and 2008 - 2010 with age group 5 always showing higher numbers compared to younger age groups. The estimated results of weight at age from the commercial samplings are also poor for the years 2002 - 2005. This can probably be explained by the comparatively low sample sizes for these years. Therefore, the use of these data as input for any age based assessment model is questionable.

## The use of survey data and application of SURBA

Probably the most suitable survey for dab is the International Beam Trawl Survey (BTS) targeting especially flat fish species in the North Sea. The problem with this survey is that it is not fully standardized and not all data are currently available via the DATRAS data portal. Furthermore, the geographical coverage is more limited compared to the International Bottom Trawl Survey (IBTS). However, the IBTS never collected any biological parameters of dab in a consistent way or covering the whole distribution range of dab. Age-length keys were available for the two Dutch BTS and the German BTS. Different options were tested to make use of age based indices in a survey based assessment model (SURBA). During the benchmark it was then agreed to use a combined beam trawl survey index for further SURBA runs which was estimated by applying the delta GAM method of Berg et al. (2014). Other index options were also tested (e.g. using BTS indices separately), but the delta GAM index showed the best internal consistency among age classes. The final SURBAR model run resulted in an overall decreasing total mortality. The spawning stock biomass and the recruitment showed increasing trends, especially from 2009 onwards. These trends could be explained by a decrease in fishing effort. Still the CPUE for the commercial fleets, but also for the surveys, increased in recent years.

## Exploratory SAM model

An exploratory assessment was done using the SAM model which is a state-space assessment model which takes the uncertainty of the assessment inputs and outputs into account. As input data for this model, the total catch weight, catch numbers at age and
catch weight at age from commercial sampling data was used. Further, stock weight at age from survey data and the combined beam trawl survey index was used. The SAM model produced similar results as the SURBA model: decreasing F, increasing stock biomass and recruitment. This is due to the fact that the SAM model puts most weight on the survey data which are the same data as used for the SURBA model. Given the unrealistic age distribution of the catch samples the fishing mortality pattern for age 5 was not realistic.

## Length based mortality estimator

The fishing mortality for dab was alternatively estimated by applying a modification of the original Gedamke-Hoenig mean length-based total mortality estimator (Gedamke and Hoenig, 2006) were total mortality Z is replaced by $\mathrm{q}^{*}$ effort +M following Then et al. (2014). Furthermore, instead of assuming constant recruitment of cohorts, a recruitment index was taken into account as proposed by a method by Gedamke et al. (2008). For this purpose a time series of effort (STECF and InterCatch data) and a recruitment index based on survey data was used. The results of the length based mortality estimator confirmed the trends in dab population dynamics which were obtained by using the SURBA model.

## Conclusion

It was agreed to keep dab a category 3.2. species according to the data limited stocks guidelines (ICES, 2012) and to use the SURBA model as future basis of advice. As input for the SURBA Dutch and German beam trawl survey data were used. A combined age based index was estimated taking into account a ship effect by applying the method by Berg et al. (2014). The weight at age data were only taken from Dutch surveys because no weight at age data were available from the German survey for most of the years. A maturity ogive was estimated by using available survey data. All data used are available in the DATRAS data base.

## References

Berg, C., Nielsen, A., Christensen, K., 2014. Evaluation of alternative age-based methods for estimating relative abundance from survey data in relation to assessment models. Fisheries Research 151: 91-99.

Gedamke, T., Hoenig, J.M., DuPaul, W.,D., Musick, J.A. (2008). Total mortality rates of the barndoor skate, Dipturus laevis, from the Gulf of Maine and Georges Bank, United States, 19632005. Fisheries Research 89: 17-25.

Gedamke, T., and J.M. Hoenig (2006). Estimating Mortality from Mean Length Data in Nonequilibrium Situations, with Application to the Assessment of Goosefish. Trans. Amer. Fish. Soc. 135:476-487.

Needle, C., 2015. Using self-testing to validate the SURBAR survey-based assessment model. Fisheries Research 171: 78-86.

Then, A., Hoenig, J. M., Hall, N.,G., Hewitt, D.,A., (2014). Evaluating the predictive performance of empirical estimators of natural mortality rate using information on over 200 fish species. ICES Journal of Marine Science 72 (1) 82-92.

## Whiting in IV and VIId (IBPWhiting-2016)

The Inter-Benchmark workshop for North Sea Whiting (IBPWhiting), chaired by Alexander Kempf, Germany for ICES and Larry Alade, USA (External Chair) met by correspondence from February to March 2016.

During the workshop it has been tested whether updated natural mortality estimates available from the ICES multi species working group (WGSAM) have a larger impact on the assessment of North Sea whiting. In addition, reference points have been reviewed in the light of new assessment results and the up-to-date ACOM guidelines have been applied to estimate potential new reference points. The precautionarity of the management strategy (fixed $\mathrm{F}=0.15$ and $15 \%$ TAC constraint) currently used as basis for advice has been tested using the same methodology (MSE simulation) as applied in 2013. The main conclusions from this work were:

1 ) The new $M$ values do not lead to an increased impact on the XSA diagnostics, but recruitment and SSB are scaled downwards because of lower $M$ values for ages 1 and 2.
2 ) Stock dynamics are driven principally by recruitment and natural mortality. FMSY does not seem to be meaningful for this stock as long as $B_{\text {loss }}$ is seen as reference point for $\mathrm{B}_{\lim }$ and $\mathrm{F}_{\mathrm{p} 05}$ has had to be used as reference point according to the ACOM guidelines. This means the management target is to avoid SSB falling below $\mathrm{B}_{\text {lim }}=\mathrm{B}_{\text {loss }}$ instead of maximizing yield. Even at low fishing mortalities, there is a significant probability to fall below $\mathrm{B}_{\text {lim }}=\mathrm{B}_{\text {loss }}$ given the current low recruitment and high natural mortalities.
3 ) The currently used management strategy (target F of 0.15 and $15 \%$ TAC constraint) is not precautionary (defined as $<5 \%$ risk to fall below $\mathrm{B}_{\mathrm{lim}}$ ) according to the MSE results. Even at a target fishing mortality of 0.1, there is a more than $5 \%$ probability to fall below $\mathrm{B}_{\mathrm{lim}}=\mathrm{B}_{\text {loss }}$ in the short and in the medium term if low to medium recruitment is assumed. As expected, this probability increases under the assumption of low recruitment only. The current management strategy has no $B_{\text {trigger }}$ value and therefore does not allow for reduced fishing mortalities to safeguard the stock. The $15 \%$ TAC constraint may further preclude fast reactions to decreasing stock abundances. The reduced flexibility for downward adjustments of TAC is problematic. In contrast, the reduced flexibility for upward adjustments is not problematic from a conservation point of view.
4 ) Eqsim gives slightly more optimistic results ( $\mathrm{F}_{\mathrm{p} 05}=0.12$ ) under current recruitment levels and no $B_{\text {trigger }}$ (fixed F exploitation independent of SBB) compared to the MSE simulations. This is caused by differences in how the recruitment is modelled. In addition, in Eqsim an assumed advice error is given as input while the MSE simulates error in the assessment and shortterm forecast.
5 ) Overall, it seems to be impossible to derive an F reference point that is precautionary under all circumstances when setting $\mathrm{B}_{\text {lim }}=\mathrm{B}_{\text {loss. }}$. Therefore, it may be best in the short term to use the standard ICES MSY advice rule (linear reduction of F below $\mathrm{B}_{\text {trigger }}$ !) and the associated $\mathrm{F}_{\mathrm{p} 05}$ of 0.15 estimated based on ACOM guidelines with Eqsim under the assumption of currently observed recruitment levels. An additional check of whether the stock is predicted in the short-term forecast to fall below $B_{\lim }$ is needed. If the stock is predicted to fall below $\mathrm{B}_{\mathrm{lim}}$, the F may be reduced further. However, whiting is fished in a mixed fishery and is considered to be a bycatch to some extent
in most North Sea demersal fisheries. There is a likely trade-off between a choke effect under the landing obligation and the target to keep the stock above $\mathrm{B}_{\mathrm{lim}}=\mathrm{B}_{\text {loss }}$ if recruitment stays at the current low level. In the longer term, the results from the inter-benchmark and the high natural mortalities estimated for this stock raise the question whether alternative management strategies (e.g. escapement strategy) are needed, and whether the currently used $B_{\lim }=B_{\text {loss }}$ (according to $A C O M$ guidelines) is realistic. In the event that environmental factors drive the stock below $B_{\text {loss }}$ despite low fishing mortalities in the near future, alternative reference points need to be explored (e.g. $B_{p a}=B_{\text {loss }}$ and $B_{\lim }=B_{p a} / 1.4$ ). A full benchmark is needed in the next two years to improve the assessment but also to evaluate any proposed alternative management strategies.

## A.1.2 Planning future benchmarks

Planning table [used for preparing the ACOM proposal of upcoming benchmarks]

| Stock | Assessment STATUS | LATEST BENCHMARK | Benchmark <br> THIS YEAR / <br> NEXT YEAR | Planning YeAR + 2 | Further PLANNING | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { cod- } \\ & 347 \mathrm{~d} \end{aligned}$ | Accepted SAM model | WKNSEA 2015 | No | No |  | Stock definition issues and area based assessments |
| $\begin{aligned} & \text { had- } \\ & 34 \end{aligned}$ | Accepted TSA model but continued exploratory assessments with SAM and SURBAR | WKHAD 2014 | IBP 2016 | No |  | May need an interbenchmark to check whether benchmark decisions are still appropiate and rules are needed to split TAC between IV, VIa and IIIa |
| $\begin{aligned} & \text { nep- } \\ & 34 \end{aligned}$ | OK | NA | No | No |  |  |
| nep-5 | Datalimited. | NA | No | No |  | TV surveys under developement |
| nep-6 | OK | 2009 WKNEPH - <br> only <br> benchmarked the <br> UWTV survey <br> process <br> 2013 benchmark | no | No |  | Fuller exploration of other input data (landings, discards, raising procedures, etc) |
| nep-7 | OK | 2009 WKNEPH - <br> only <br> benchmarked the <br> UWTV survey <br> process <br> 2013 benchmark | No | No |  | Fuller exploration of other input data (landings, discards, raising procedures, etc) |
| nep-8 | OK | 2009 WKNEPH - <br> only <br> benchmarked the <br> UWTV survey <br> process <br> 2013 benchmark | No | No |  | Fuller exploration of other input data (landings, discards, raising procedures, etc) |
| nep-9 | OK | 2009 WKNEPH - <br> only <br> benchmarked the <br> UWTV survey <br> process <br> 2013 benchmark | No | No |  | Fuller exploration of other input data (landings, discards, raising procedures, etc) |
| $\begin{aligned} & \text { nep- } \\ & 10 \end{aligned}$ | Data limited |  | No | No |  |  |


| Stock | Assessment STATUS | LATEST BENCHMARK | Benchmark THIS YEAR / NEXT YEAR | Planning YEAR +2 | Further <br> PLANNING | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { nep- } \\ & 32 \end{aligned}$ | Data limited | 2013 | 2016 | No |  | Exploration of all available data, incl new Norw electronic logbooks |
| $\begin{aligned} & \text { nep- } \\ & 33 \end{aligned}$ | Data limited | No benchmark ever on this stock, mainly due to lack of data | no | No |  | More data should be made available for this stock before a new benchmark |
| $\begin{aligned} & \text { nep- } \\ & \text { IIIa } \end{aligned}$ | OK | No benchmark ever on this stock, 2009 WKNEPH - only benchmarked the UWTV survey process. | 2016 | No |  | Fuller <br> exploration of other input data (landings, discards, raising procedures, etc) |
| $\begin{aligned} & \text { nep- } \\ & 34 \end{aligned}$ | Data limited |  | no | no |  | Fuller exploration of other input data (landings, discards, raising procedures, etc) |
| $\begin{aligned} & \text { nop- } \\ & 34 \end{aligned}$ | OK | 2012 | 2016 | No |  | Assessment method, Reference levels in ecosystem / multi-species context, commercial fishyer tuning time series, average recruitment used in forecast for forecast year. |
| ple- <br> eche | ok | WKPLE 2015 | No | No |  | Stock definition and mixing with North Sea stock still unclear |


| Stock | Assessment status | LATEST BENCHMARK | Benchmark <br> this Year / NEXT YEAR | Planning Year + 2 | Further <br> PLANNING | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ple- <br> nsea | OK | 2009, Interbenchmark procedure in 2013 | 2017 | No | - | Changes in catchability for indices of young ages (13) may need to be addressed again in a future benchmark. Potential removal of the SNS index now that the two BTS indices (Isis and Tridens) have been combined. Possible inclusion of the IBTS survey. |
| sai- 3a46 | Accepted FLXSA assessment but exploratory assessments with SAM. | WKNSEA 2016 | no | no |  | stock identity, discards data, revised CPUE indices, investigation of research surveys (whether they are adequate for use as tuning indices), investigate the lack of 3-year olds in the Q3 survey (fish not appearing until age 4). Test alternative models |
| sol- <br> eche | OK | 2009 | 2017 | no |  | Evaluating available Tuning series and alternative models |
| sol- <br> nsea | OK | WKNSEA 2015 | no | no |  | Inclusion of Belgium BTS and new CPUE time series when available. |


| Stock | Assessment STATUS | LATEST BENCHMARK | Benchmark this Year / NEXT YEAR | Planning <br> Year +2 | Further <br> PLANNING | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| tur- <br> nsea | In development | IBP 2015 | no | no | . | Still hihgly uncertain assessment. More data and better survey information needed |
| whg47d | Update deviating from benchmark | $\begin{aligned} & \text { 2009, } 2013 \text { IBP } \\ & 2016 \end{aligned}$ | no | 2018 |  | Change in catchability of young fish in IBTS surveys requires application of different but extant method. New natural mortalites tested in IBP2016 |
| whgkask | Data limited stock | No | $\begin{aligned} & \text { IBP } \\ & 2016 / 17 \end{aligned}$ | No |  | Aim to move from Cat 5 to Cat 3 stock by developing biomass index, review historical catch data and run SPiCT-model |
| Polnsea | Data limited stock | no | no | no |  |  |
| $\begin{aligned} & \text { mur- } \\ & 347 \mathrm{~d} \end{aligned}$ | quantitative <br> advice for <br> data-limited <br> stocks | WKNSEA 2015 | no | no |  | Length based assessment methods should be further investigated |
| bllnsea | Data-limited | No/WGNEW | no | no |  |  |
| dab- <br> nsea | Data-limited | WKNSEA 2016 | no | no |  | Test analytical assessment methods |
| Fle- <br> nsea | Data-limited | No/WGNEW | no | 2018 |  | Collate length and age based data, test analytical assessment methods. |
| Lemnsea | Data-limited | No/WGNEW | no | 2018 |  | Collate length and age based data, test analytical assessment methods. |


| Stock | Assessment STATUS | LATEST BENCHMARK | Benchmark <br> this Year / NEXT YEAR | Planning Year + 2 | Further <br> PLANNING | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Witnsea | Data-limited | No/WGNEW | no | 2018 |  | Test analytical assessment methods |
| Turkask | Data-limited | No/WGNEW | No | no | 2019 | IBP may be needed to check current stock definition |
| GUR | Data-limited | No/WGNEW | No | no |  |  |

## A.1.3 Issue lists, workplans and progress reports for stocks with upcoming benchmarks

Benchmarks 2016
Apart from the benchmarks for Norway Pout and Nephrops in IIIa and 32, detailed below, interbenchmarks are planned for Northern Shelf haddock in the summer of 2016 to deal with the retrospective pattern problem that caused a delay in the assessment and advice for this stock, and for whiting in 3.a prior to WGNSSK in 2017 to attempt to upgrade the assessment to a category 3 stock.

## Norway pout

Terms of Reference were issued by ACOM, as follows:
2015/2/ACOM36 A Benchmark Workshop on Norway Pout (WKPout), chaired by External Chair XX, xx and ICES Chair José De Oliveira, UK, and attended by two invited external experts Jerry Ault, USA, Verena Trenkel, France and Daniel Hennen, USA will be established 18-19 May 2016 for a data evaluation meeting at ICES Headquarters and in tbc., 22-26 August 2016 for a Benchmark meeting, to:
a ) Evaluate the appropriateness of data and methods to determine stock status and investigate methods for short term outlook taking agreed or proposed management plans into account for the stocks listed in the text table below. The evaluation shall include consideration of:
i) Stock identity and migration issues;
ii ) Life-history data;
iii ) Fishery-dependent and fishery-independent data;
iv ) Further inclusion of environmental drivers, multi-species information, and ecosystem impacts for stock dynamics in the assessments and outlook
b ) Agree and document the preferred method for evaluating stock status and (where applicable) short term forecast and update the stock annex as appropriate. Knowledge about environmental drivers, including multispecies interactions, and ecosystem impacts should be integrated in the methodology

If no analytical assessment method can be agreed, then an alternative method (the former method, or following the ICES approach for stocks without analytical assessments) should be put forward;
c ) Re-examine and update, if necessary, MSY and PA reference points according to ICES guidelines (see reports of WKMSYREF 2-4 and the Technical document on reference points);
d ) Bearing in mind that the catch advice for Norway pout is for the period November 1 of the assessment year to October 31 of the following year, evaluate the settings of the escapement strategy used in the ICES MSY approach and the value of $\mathrm{F}_{\text {cap. }}$
e ) Develop recommendations for future improving of the assessment methodology and data collection;
f) As part of the evaluation:
i) Conduct a 3 day data evaluation workshop. Stakeholders are invited to contribute data (including data from non-traditional sources) and to contribute to data preparation and evaluation of data quality. As part of
the data evaluation workshop consider the quality of data including discard and estimates of misreporting of landings;
ii ) Following the data evaluation correspondence work, produce working documents to be reviewed during the Benchmark meeting; these documents should be available at least 7 days prior to the meeting

| Stocks |  | STOCK LEADER |
| ---: | :--- | :--- |
| Nop-34 | J. Rasmus Nielsen |  |

The Benchmark Workshop will report by 15 September 2016 for the attention of ACOM.

## Proposed working papers/analyses

1 ) Seasonal SAM assessment development, including 3-5-year back comparison with SXSA, and sensitivity analysis [led by Rasmus Nielsen]

2 ) Forecast methodology, and MSY and PA reference points, including implications of sensitivity analysis and ecosystem considerations [led by Rasmus Nielsen]
3 ) Norwegian fishery description and CPUE analysis [led by Espen Johnsen]
4 ) Danish fishery description and CPUE analysis [led by Rasmus Nielsen]
5 ) Existing supporting documents on historic catches (Industrial fisheries Expert Group, EU-Norway negotiations) [co-ordinated by Rasmus Nielsen]
6 ) Age-reading data comparisons [co-ordinated by Rasmus Nielsen]
7 ) Norwegian Shrimp survey estimates east of Norwegian trench (comparing perceptions with IBTS surveys), as supplementary information only (not as a tuning series) [led by Espen Johnsen]

## Workplan (recommendations from DCWKPOUT)

- Keep current data sets (current catch, current historic CPUE and current IBTS Q1, Q3 and UK-Q3-EGFS and UK-Q3-SGFS) and develop SAM assessment
- Check whether discards can be included or not (include if yes)
- Keep SXSA assessment as an alternative assessment for comparison at the benchmark
- Develop forecast methodology
- Develop MSY and PA reference points (including ecosystem considerations)
- Conduct sensitivity analyses:
- Remove historic commercial CPUE from the assessment
- Compile time series of effort and compare to changes in F
- Norwegian and Danish data available

| Stock | Nop-34 |  |
| :--- | :--- | :--- |
| Benchmark | Year: 2016 |  |
| Stock coordinator | Name: J. Rasmus Nielsen | Email: rn@aqua.dtu.dk |
| Stock assessor | Name: J. Rasmus Nielsen | Email: rn@aqua.dtu.dk |
| Data contact | Name: J. Rasmus Nielsen, Espen <br> Jonssen | Email: rn@aqua.dtu.dk |

## Nephrops in IIIa

## Data needed

Genetic information

- VMS information
- Sediment type maps
- Landings data
- Discard data
- Growth data
- Size and sex distribution
- Maturity information
- Length-weight data
- UWTV survey data


## Current assessment issues

- Redefine Nephrops' spatial distribution in IIIa and update population estimates retrospectively based on the new area estimate.
- Create UWTV survey reference footage
- Update Length Cohort Analysis (LCA)


## Proposed working papers/analyses

- The current spatial distribution of Nephrops grounds is based on Danish and Swedish VMS data from 2008-2010. The boarders to the grounds have been arbitrarily defined using two different methods. The most recent VMS data (2013-2015) together with Swedish logbook data from creel and $<12 \mathrm{~m}$ vessels and sediment maps will be combined to establish a new distribution of Nephrops suitable habitats. The new area estimates will be used to update the Nephrops stock population time series.
- Following the same protocol for other FU's where UWTV surveys exist, reference footage will be compiled, comprising footage of poor, average and high quality as well as different burrow densities (low, medium and high density). A manual will complement the footage, describing how to use the footage.
- The current LCA is based on length frequency data from 2008-2010 and assumes a discard survival of $25 \%$. Under a landing obligation, survivability will be zero as all individuals are to be landed. Therefore, the LCA needs to be updated with new data as well as different survival rates ( 0,25 and $50 \%$ ). There are also discussions around lowering the MLS (MCRS) for Nephrops in the Skagerrak and Kattegat from 40 mm carapace length to 30 mm . This would change the discard pattern in the fisheries and would also require an update to the LCA. Therefore, the sensitivity of the LCA to different MLSs will be assessed.


## Workplan

- Redefine spatial distribution. (Jordan - November/December 2015 WGNEPS)
- Create UWTV survey reference footage (Mats - November/December 2015 - WGNEPS)
- Update Length Cohort Analysis (LCA) (Helen Dobby/Mats - February 2016)
- 

| STоск | Nep IIIA |  |
| :--- | :--- | :--- |
| Stock <br> coordinator | Name: Mats Ulmestrand | Email: mats.ulmestrand |
| Stock assessor | Name: Jordan P. Feekings | Email: jpfe@dtu.aqua.dk |
| Data contact | Name: | Email: |


| Issue | Problem/Aim | Work needed / possible direction of solution | Data needed to be able to DO THIS: ARE THESE available / Where should these come FROM? | External expertise <br> needed at <br> benchmark <br> TYPE OF EXPERTISE / PROPOSED NAMES |
| :---: | :---: | :---: | :---: | :---: |
| (New) data to be considered and/or quantified1 | Redefine area of all subareas using up to date information. | Compile available data sources from all available years. | DK \& SW VMS data, SW logbook data from < 12 m vessels and creel vessels, sediment maps etc. Data are available. |  |
|  | Create UWTV reference footage. | Review selection of UWTV footage and establish consensus counts | UWTV survey footage. Data are available |  |
| Tuning series | Standardize Swedish lpue. |  | Logbook data are available. |  |
| Discards | Effects of changing the MLS <br> Effects of a landing obligation | Bio-economic analysis of changing the MLS for Nephrops. | VMS, <br> landings, <br> discards, stock <br> estimates, <br> price data, <br> biological data <br> (e.g. size and <br> sex <br> distribution, <br> female and <br> male <br> maturity). |  |

${ }^{1}$ Include all issues that you think may be relevant, even if you do not have the specific expertise at hand. If need be, the Secretariat will facilitate finding the necessary expertise to fill in the topic. There may be items in this list that result in 'action points for future work' rather than being implemented in the assessment in one benchmark.

| Issue | Problem/Aim | Work needed / POSSIBLE DIRECTION OF SOLUTION | Data needed to be Able to DO THIS: ARE THESE AVAILABLE / WHERE SHOULD THESE COME FROM? | External expertise <br> NEEDED AT <br> benchmark <br> TYPE OF EXPERTISE / PROPOSED NAMES |
| :---: | :---: | :---: | :---: | :---: |
| Biological <br> Parameters | Growth parameter update <br> Length-weight update |  |  |  |
| Assessment method | The UWTV survey method is not possible for creel areas. Possible inclusion of a length based assessment model to compliment the UWTV survey. | Develop and finalize a length based model for Nephrops. | Catch data. <br> Data are available for trawled and creeled areas. | Anders Nielsen, DTU Aqua |
| Biological <br> Reference <br> Points |  | Model work should provide these reference points. | Proxies for <br> Fmsy exist from LCA. <br> Btrigger estimate requires a longer UWTV time series. |  |

## Nephrops 32

## Data needed

- Danish data from at-sea-observers (discard, lfd, sex ratio)
- Norwegian shrimp survey data
- Norwegian electronic log book data
- Norwegian data from recreational Nephrops fishery
- Norwegian Coast Guard data from vessel inspections


## Current assessment issues

## Danish data

- Investigate possibilities for obtaining biological data (maturity, weight, length) from the Danish at-sea-observer program
- Analyze discard data (strange values in recent years), document old and new sampling procedures, agree on standard sampling procedure

Norwegian data

- Explore possibilities for obtaining a Nephrops biomass index from the survey data
- Analyze biological data from recent studies on recreational fishery along Norwegian coast (sex ratio, length)
- Analyze so far unused total length data (TL) from Coast Guard inspections
- Investigate possibilities for obtaining discard data from Coast Guard inspections
- Explore the area calculations used for estimating harvest rates. Is the current estimated area too large?
- Explore electronic logbook data from 2011 onwards and establish new LPUE time series from respectively shrimp and Nephrops trawls.


## Proposed working papers/analyses

Working paper 1: Danish at-sea-observer program: documentation of procedures and possibilities for obtaining biological data.

Working paper 2: New biomass index time series from Norwegian electronic logbook data.

Working paper 3: A new biomass index from the Norwegian annual shrimp survey in the Norwegian Deep?

Working paper 4: Exploration of Norwegian Coast Guard inspections data: discard and length frequency distributions.

Working paper 5: Biological data from recent studies of the Norwegian recreational fishery - does the coastal Nephrops differ from animals on offshore fishing grounds?

## Workplan

Most of the work will be done autumn 2015, before the data workshop. Jordan Feekings will work on the Danish discard data and participate in discussions on the area estimates. Guldborg Søvik will analyze the Norwegian data, with the aid of colleagues at IMR.

| Sтоск | NePHROPS IN FU 32 |  |
| :--- | :--- | :--- |
| Stock <br> coordinator | Name: Guldborg Søvik | Email: guldborg.soevik@imr.no |
| Stock <br> assessor | Name: Guldborg Søvik | Email: guldborg.soevik@imr.no |
| Data contact | Name: Guldborg Søvik | Email: guldborg.soevik@imr.no |


| Issue | Problem/Aim | Work needed / possible direction of solution | Data needed to be Able to do THIS: ARE THESE available / Where should these come FROM? | EXternal expertise needed at benchmark TYPE OF EXPERTISE / proposed names |
| :---: | :---: | :---: | :---: | :---: |
| New data to be considered | Current stock definition is based on spatial distribution. The level of larval exchange with Skagerrak is unknown. | Work is ongoing at IMR on collecting/analysing new genetic data from Kattegat, Skagerrak and the Norwegian Deep, as part of a Scandinavian Interreg project | New information will be available by end of 2014/beginning of 2015 | Relevant <br> Nephrops experts. <br> The benchmark should be arranged together with the planned benchmark for FU 3 and 4 |
| New analyses to be considered | At WKNEPH 2013 it was suggested that the Nephrops data in the Norwegian annual shrimp survey should be analysed in order to determine if they can be used to provide a fishery independent biomass index | Analyse annual Nephrops catches from shrimp survey | Nephrops data from Norwegian shrimp survey exist back to 1997 | Time series experts |
| Biological <br> Parameters | No biological data exist for this stock | Collection of biological data from stock component along the Norwegian coast | Collection of data from the Norwegian recreational fishery was initiated in 2012. It is expected that by 2016 data on sex ratio, length frequency distribution and life cycle will be available | relevant Nephrops experts |


| Issue | Problem/Aim | Work needed / POSSIBLE DIRECTION of solution | Data needed to be able to do this: ARE THESE available / where should these come FROM? | External expertise <br> needed at <br> benchmark <br> TYPE OF EXPERTISE / <br> PROPOSED NAMES |
| :---: | :---: | :---: | :---: | :---: |
| Biological <br> Parameters | No biological data exist for this stock | Collection of biological data from stock component along the western part of the Norwegian Deep | Investigate possibilities for obtaining biological data from the Danish at-sea-samplingprogramme | relevant Nephrops experts |
| Fishery patterns | How much does spatial and temporal variation in the fishery influence stock biomass indices (LPUE)? | The Danish fishery (1990s until present) has gone through large changes. Spatial and temporal patterns in the fishery need to be explored. | Annual Danish VMS data exist | relevant Nephrops experts |

## Benchmarks 2017

## Plaice in IV and IIIaN

## Data needed

Natural mortality data
Maturity data
Landings and discards data per fleet
Tuning series of plaice from IBTS, SNS, BTS-Tridens, and BTS-ISIS

Current assessment and forecast issues
XSA

Proposed working papers/analyses
TBA

## Workplan

TBA

| Sтоск | Plaice-North Sea Skagerrak |  |
| :---: | :---: | :---: |
| Stock coordinator | Name: Chun Chen/Jan Jaap Poos/Ruben Verkempynck | Email: chun.chen@wur.nl/janjaap.poos@wur.nl |
| Stock assessor | Name: Chun Chen | Email: chun.chen@wur.nl |
| Data contact | Name: Chun Chen/Ruben Verkempynck | Email: chun.chen@wur.nl/ruben.verkempynck@wur.nl |



| Issue | Problem/Aim | Work needed / POSSIBLE DIRECTION of solution | Data needed to be able to DO THIS: ARE THESE available / WHERE SHOULD these come fROM? | External expertise <br> needed at <br> benchmark <br> TYPE OF EXPERTISE / proposed names |
| :---: | :---: | :---: | :---: | :---: |
| Tuning series | Exclusion of SNS or combine SNS with BTS | Make a combined index SNS/BTS and explore assessments with or without SNS and combined index | SNS, BTS <br> indices | Chun Chen |
| Tuning series | Inclusion of IBTS as tuning fleet | IBTS index | IBTS index | Clara Ulrich, Chun Chen, Casper Berg |
| Discards |  |  |  |  |
| Biological <br> Parameters | Natural mortality Maturity | Review of basis for natural mortality. Literature review, model estimates of M Review of basis for maturity. Literature review, model estimates of maturity |  | Chun Chen, Jan Jaap Poos, Ruben Verkempynck |
| Assessment method | Explore other models (SCAA, SAM, Aarts and Poos). |  |  | Chun Chen, Jan Jaap Poos, Ruben Verkempynck |
| Biological <br> Reference <br> Points | Revision Fmsy and MSYBtrigger after inclusion of Skagerrak |  |  | Chun Chen, Jan Jaap Poos, Ruben Verkempynck |
| Management plan | Revision of the North Sea management plan after adding SK AND <br> Implementation of stage 2 of the MP |  |  | Chun Chen, Jan Jaap Poos, Ruben Verkempynck |

## Sole in VIId

## Data needed

- Discard time series from BEL, FRA and UK
- Belgian discard data at quarterly scale
- Alternative fishery-independent recruitment data
- National commercial CPUE indices
- Maturity at age/length data not already included in DATRAS
- Any tagging data/information or genetic analysis


## Current assessment issues

See Table below.

## Proposed working papers/analyses

Working paper on maturity (Lies Vansteenbrugge, Bart Vanelslander, Sofie Nimmegeers)

## Workplan

- Discard time series (at quarterly scale) ready for the Data Evaluation Workshop
- Alternative fishery-independent recruitment data and national commercial CPUE indices ready for the Data Evaluation Workshop
- Alternative maturity ogive and corresponding working paper ready by the actual benchmark
- Comparative runs using different assessment models (XSA, AAP and SAM) at the benchmark
- Simulation model runs to calculate reference points (especially Blim and Fbar) at the benchmark
- 


## Other working groups to be involved

Not applicable

| Sтоск | SoL-есне |  |
| :--- | :--- | :--- |
| Stock <br> coordinator | Name: Lies Vansteenbrugge | Email: lies.vansteenbrugge@ilvo.vlaanderen.be |
| Stock <br> assessor | Name: Lies Vansteenbrugge <br> and Bart Vanelslander | Email: $\underline{\text { lies.vansteenbrugge@ilvo.vlaanderen.be }}$ <br> bart.vanelslander@ilvo.vlaanderen.be |
| Data contact | Name: Bart Vanelslander | Email: bart.vanelslander@ilvo.vlaanderen.be |


| Issue | Problem/Aim | Work needed / POSSIBLE DIRECTION OF solution | Data needed to be ABLE TO DO THIS: ARE THESE avallable / where Should these COME FROM? | External expertise needed at <br> benchmark <br> TYPE OF EXPERTISE / PROPOSED NAMES |
| :---: | :---: | :---: | :---: | :---: |
| (New) data to be considered and/or quantified2 | Additional M predator relations | Not at the moment |  |  |
|  | Prey relations | Not at the moment |  |  |
|  | Ecosystem drivers | Not at the moment |  |  |
|  | Other ecosystem parameters that may need to be explored? | Not at the moment |  |  |
| Tuning series | As the UK Young Fish Survey (YFS) stopped in 2006, the French YFS and the UK(E\&W)-BTS-Q3 are the two remaining fisheryindependent sources of information on age 1. There is however doubt around to what extent the BTS, that fishes with a mesh size of 40 mm in the codend, catches 1 year old sole in a quantitative way (BTS tuning series used in other assessments often start at age 2). <br> Furthermore, the French YFS is probably not providing the correct recruitment estimates as it only covers part of VIId (potential impact on forecast). | Analysis: test runs with UK(E\&W)-BTS-Q3 for ages 2 and older only. New data : Long-term : UK to consider picking up the YFS in VIId again, or FRA to consider extending the French YFS into UK waters. <br> Short-term : FRA/UK to check for other data sources that could inform WKNSEA <br> 2017 on the strength of incoming year classes (especially ages 1-2), so it can be evaluated to what extent | Data are delivered annually by CEFAS. <br> If the UK-YFS is reinstated, this will not lead to a time series that will already be useful at the time of WKNSEA 2017. <br> FRA and UK | Potentially people from FRA and/or UK to present new data. <br> No specific extra expertise needed for analysis of these data. |

${ }^{2}$ Include all issues that you think may be relevant, even if you do not have the specific expertise at hand. If need be, the Secretariat will facilitate finding the necessary expertise to fill in the topic. There may be items in this list that result in 'action points for future work' rather than being implemented in the assessment in one benchmark.

| Issue | Problem/Aim | Work needed / POSSIBLE DIRECTION OF SOLUTION | DATA NEEDED TO BE ABLE TO DO THIS: ARE THESE <br> AVAILABLE / WHERE SHOULD THESE COME FROM? | External expertise <br> NEEDED AT <br> benchmark <br> TYPE OF EXPERTISE / <br> PROPOSED NAMES |
| :---: | :---: | :---: | :---: | :---: |
|  |  | these data correspond to the views presented by the French YFS. |  |  |
| Discards | Discards are not included in the assessment. | BEL, FRA and UK to compose time series of raised discard data (as long as possible). | Data available from Belgian, French and English discard sampling programmes. | No specific expertise needed from countries delivering the discard time series. |
|  | FRA and UK upload quarterly discard data for all sampled métiers to Intercatch, whereas Belgium only uploads annual estimates. As <br> Belgium only samples the métier TBB_DEF_70-99, a métier with a high share in the total effort/landings of soleche and that is generally characterised by high discard rates and discard patterns that can be very different compared to the other métiers that are predominant in the FRA and UK sampling programmes, it would be desirable to also obtain quarterly discard estimates from Belgium for this metier. | Belgium to compute quarterly discard estimates for sol-eche (TBB_DEF_70- <br> 99) based on the available data (some countries have a comparable sampling level, but deliver quarterly data with a low qualitative selfevaluation, where Belgium chooses to only compute annual estimates with a higher selfevaluation) | Data available in Belgium (ILVO), but need to be processed differently and delivered so the impact on the assessment can be evaluated. <br> Not relevant to WKNSEA 2017. | ILVO (Lies <br> Vansteenbrugge, <br> Bart Vanelslander, <br> Sofie Nimmegeers) |
| Biological <br> Parameters | A knife-edged maturity ogive, with full maturation from age 3 onwards is used in the assessment. No new data have been explored for a long time. | Investigate all available trawl survey maturity data to come up with a maturity ogive that is supported by recent data. | Data available in DATRAS. | ILVO (Lies <br> Vansteenbrugge, <br> Bart Vanelslander, <br> Sofie Nimmegeers) |


| Issue | Problem/Aim | Work needed / POSSIBLE DIRECTION OF SOLUTION | Data needed to be ABLE TO DO THIS: ARE THESE AVAILABLE / WHERE SHOULD THESE COME FROM? | External expertise <br> NEEDED AT <br> BENCHMARK <br> TYPE OF EXPERTISE / PROPOSED NAMES |
| :---: | :---: | :---: | :---: | :---: |
| Assessment method | The current XSA behaves well, but when discard time series get included as input data, other models need to be tested: AAP, potentially SAM. | Carry out comparative runs using different models. | / | Experts in running AAP, SAM. This expertise is currently not available at ILVOBelgium. |
| Biological <br> Reference <br> Points | Revision of reference points. Preliminary analyses during WGNSSK 2014 already suggested that a revision of Fmsy could be considered. No Blim is identified and Fbar is calculated on other ages than Fmsy is. | Computation of potential new reference points. | Data available from assessment. | Experts in computation of reference points. This expertise is currently not available at ILVOBelgium. |

## Proposed Benchmarks 2018

## Whiting in 4 and 7.d

## Data needed

- Catch (landings, discards, IBC) data back to 2002 (BEL, DNK, FR, GER, NDL, NOR, SWE, UK ENG, UK SCO).
- Biological data available from commercial sampling and IBTS surveys (including (maturity, length distributions, age distributions, individual weights, sex ratios)
- IBTS survey data


## Current assessment and forecast

Currently, assessment is done using a FLXSA, assuming catches to be exact. And for comparison a SURBAR analysis is run. The IBTS survey indices quarter 1 and 3 are used in the analysis. Advice is given on ICES area Subdivision IV and Division VIId combined. The advice per area is then split between areas based on the landings ratios. The TAC is given for Subdivision IV separately, and for Division VIId in combination with VIIb-k.

## Proposed analysis

A more detailed stock structure has been proposed dividing a northern North Sea stock from a southern North Sea stock (Holmes et al 2014, SGSIMUW 2005). It needs to be determined whether differences in biological characteristics, such as growth maturity, exist between suggested new substocks. Also, for the analysis it would be necessary to determine whether available historical landings and discard data is sufficient for a subdivision into new assessment areas.

Maturity values at age used in the XSA originate from analysis from the 1980ies. A check is suggested whether an update of the used values is necessary, also with regard to new stock units. National catch sampling and survey data may give those information.

Generally, an update of the assessment model is suggested. Available models include SAM or TSA. This would address issues with variability in catches, and catchability changes.

Within the framework of a benchmark the choice of input data into a short term forecast can be addressed (individual weights at age, recruitment estimates, fishing mortality at age estimates).

Holmes, S.J., Millar, C.P., Fryer, R.J., Wright, P.J., 2014. Gadoid dynamics: differing perceptions when contrasting stock vs. population trends and its implications to management. ICES J Mar Sci. 71: 1433-1442.

ICES, 2005. Report of the study group on stock identity and management units of whiting (SGSIMUW), 15-17 March 2005, Aberdeen, UK. ICES CM 2005/G:03: 50pp.

## Workplan

Compilation of Intercatch data for recent years of Landings, Discards and IBC
Compilation of IBTS survey data including biological information
Checking whether data with resolution based on stock structure is available

Exploratory assessment runs
(WG to be involved: IBTS WG, WGBIOP)

| Sтоск | WhG47d |  |
| :--- | :--- | :--- |
| Stock coordinator | Name: Tanja Miethe | Email: tanja.miethe@marlab.ac.uk |
| Stock assessor | Name: Tanja Miethe | Email: tanja.miethe@marlab.ac.uk |
| Data contact | Name: Tanja Miethe | Email: tanja.miethe@marlab.ac.uk |


| Issue | Problem/Aim | Work needed / POSSIBLE DIRECTION OF SOLUTION | Data needed to be able TO DO THIS: ARE THESE AVAILABLE / WHERE should these come FROM? | External EXPERTISE NEEDED AT benchmark TYPE OF EXPERTISE / PROPOSED NAMES |
| :---: | :---: | :---: | :---: | :---: |
| (New) data to be | Additional Mpredator relations |  |  |  |
|  | Prey relations |  |  |  |
| Considered <br> and/or <br> quantified3 | Ecosystem drivers |  |  |  |
|  | Other ecosystem parameters that may need to be explored? |  |  |  |
| Tuning series |  |  |  |  |
| Discards |  |  |  |  |
| Biological <br> Parameters | Maturity <br> (differences for areas, stock units) | Compile available data on maturity | IBTS Survey data (DATRAS), commercial sampling data | Coby Needle, <br> Peter Wright <br> (Scotland) |
|  | Growth (differences for areas, stock units) | Compile available data on growth (length at age) from surveys | IBTS Survey data, commercial sampling data | Coby Needle, Peter Wright (Scotland) |
| Assessment method | XSA treats catch data as exact | Develop and test new assessment model (e.g. SAM, TSA) | Catch data, survey data | Anders Nielsen (DTUAqua) Rob Fryer (Scotland) |
|  | Stock structure | Compile available data on catches by area or stock units | Catch data (landings, discards, IBC) | Liz Clarke (Scotland) |


| Issue | Problem/Aim | Work needed / possible DIRECTION OF solution | Data needed to be able TO DO THIS: ARE THESE available / where should these come FROM? | External EXPERTISE needed at benchmark <br> TYPE OF EXPERTISE / PROPOSED NAMES |
| :---: | :---: | :---: | :---: | :---: |
| Biological <br> Reference <br> Points |  |  |  |  |
| Short term forecast update | Check choice of input data to STF |  | Catch data(landings, discards, ibc), survey outputs |  |

## Witch in IIIa, IV and VIId

Witch flounder (Glyptocephalus cynoglossus) in Subarea IV, Division IIIa and VIId is one of the stocks that were classified as category 3 stocks following the guidelines of the ICES Data Limited during WGNEW 2013.

In 2009 witch flounder has been included as a mandatory species in the EU Data Collection Framework (DCF). Accordingly, Denmark and Sweden started the regular sampling of biological data, i.e. length, weight, maturity status and age, in IIIa and IV both in discards and landings. Scotland has also been collecting biological samples since 2009 but only from the landings.

Abundance indices for witch flounder show a declining trend since the peak observed in 2000, followed by an increase in the most recent years both in landings and survey data. For this stock an exploratory Extended Survivors Analysis (XSA) was performed in 2013 which indicate that fishing mortality is above potential FMSY proxies (ICES WGNEW, 2013).

The model used in 2013 could be improved during the benchmark, given the inclusion of additional years and the availability of discards data allowing the use of catch data instead of only landings.

## Tuning indices

In the XSA run in 2013 data, from IBTS Q1 were used as tuning index. During the benchmark those survey data will be updated and investigated by area, in order to check whether one of the areas, namely IIIa showing a larger number of fish caught per hour, is mostly driving the observed pattern. Concerning IBTS data Q3, data were not used in the previous assessment attempt, given the shorter time series. However, the abundance trend observed in Q3 did not reflect the one in Q1 possibly due to different spatial distribution of this species during the two periods. Given the scarce knowledge about the biological demographic dynamics of this stock it is not possible to define with certainty whether the observed difference is due to differential biological (feeding and reproductive) phases.

It was noticed that there is some difference in the coverage between the two surveys thus some statistical rectangles are not covered in one of the quarters. The next step is thus to consider a standardized area to ensure that the observed trends are mirroring the abundance of the same statistical rectangles and thus the detected difference is not due to a diverse coverage. The possibility of using Beam Trawl Survey (BTS) data in the same standardized area will also be investigated.

## Assessment models to be investigated

The possibility of using a production model was discussed but the availability of discards data only from 2002 makes the use of this kind of model inappropriate, as a longer time series of catches is needed. Discards data will be used for the first time, mainly based on Swedish and Danish data, for running an updated XSA using catch data instead of only landings and an additional year. An a4a (Assessment for All Initiative) together with an SS3 will be also tested.

## Working documents

Three working papers will be prepared:
1 ) Description of landings and discards raising procedure
2 ) Exploration of survey data to be used as tuning index
3 ) Assessment models trials

| Stock | Witch IIIA, IV AND VIId |  |
| :--- | :--- | :--- |
| Stock <br> coordinator | Name: Francesca Vitale | Email: francesca.vitale@slu.se |
| Stock assessor | Name: Max Cardinale | Email: <br> massimiliano.cardinale@slu.se |
| Data contact | Name: Francesca Vitale | Email: francesca.vitale@slu.se |


| Issue | Problem/Aim | Work needed / POSSIBLE DIRECTION OF SOLUTION | Data needed to be Able to DO THIS: ARE THESE AVAILABLE / WHERE SHOULD THESE COME FROM? | EXTERNAL EXPERTISE <br> needed at <br> BENCHMARK <br> TYPE OF EXPERTISE / PROPOSED NAMES |
| :---: | :---: | :---: | :---: | :---: |
| (New) data to be | Additional M predator relations | Not at the moment |  |  |
| Considered and/or quantified4 | Prey relations | Not at the moment |  |  |
|  | Ecosystem drivers | Not at the moment |  |  |
|  | Other ecosystem parameters that may need to be explored? | Not at the moment |  |  |
| Tuning series | IBTS Q1 and Q3, BTS Q1 and Q3 | The series are available and need just to be updated | DATRAS | None |
| Discards | Partially available on Intercatch only for Sweden, Netherland and Denmark in 2013 | MS to submit discards information for the rest of the time series | Estimation of discards by country and by area | None |
| Biological <br> Parameters survey | MO, WAA, NM | The series are available and need to be updated. | SLU AQUA will collate and update the biological data | None |


|  | Ongoing <br> maturity studies. |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Biological <br> Parameters <br> catch | MS to submit <br> landings <br> information <br> (number at age <br> and weight at <br> age) for the rest <br> of the time series | SLU AQUA | will collate <br> and compile <br> the biological <br> data |  |

## Progress towards benchmark

Tuning series: IBTS time series have been updated including 2016
Biological paramters, survey: Determination of maturity schedule and ageing technique almost finalised

Biological paramters, catch: Catch data submitted by MSs from 2009
Assessment method: To be conducted in 2016/2017
Biological reference points: To be conducted after the benchmark

## Lemon sole in Subarea 4 and Divisions 3.a and 7.d

## Data needed

- Catch (landings and discards) data back to 2002 (BEL, NDL, GER, DNK, SWE, NOR, UK ENG, UK SCO, FRA).
- Effort data by single métiers as potential additional source of information to estimate the amount of discards where no sufficient information from sampling programs is available.
- Available IBTS and BTS survey data on lemon sole.
- Available biological data from scientific surveys and sampling programs of the commercial fleets (length distributions, age distributions, spatial distributions, weight, maturity, sex ratios).


## Current assessment issues

Currently lemon sole is treated as a data limited species and the stock perception is derived from simple survey trends and catch data. For the current assessment method the IBTS-Q1 index (mature biomass in kg ) is used as basis for the trend based analysis (DLS 3.2. method). The IBTS surveys appear to cover the stock distribution well, but may not provide representative indices as the GOV gear is not optimized for flatfish sampling. The available beam-trawl surveys should provide a more representative index, but may not cover the full stock distribution. Survey data needs to be analysed to see if these issues can be circumvented. Further, it was often argued within the WGNSSK that the indices used for the DLS 3.2. method do not come along with uncertainty estimates and that these estimates should be calculated.

For commercial catches to be used in assessments, discards need to be quantified for lemon sole. Data are available to do this for 2012-2015, but more discard data sampled under the DCF should be available and should be uploaded into the InterCatch data portal by all relevant institutes.

Although lemon sole is treated as a data-limited stock, much data exist which are not utilized today. The available surveys can provide biological data for a number of years (including lengths, ages, maturity, weights and spatial location), and national catchsampling programmes may be able to provide further valuable information. Agebased stock abundance indices can be generated from survey data, and could be used as the basis for survey-based assessment methods such as SURBAR. The distribution of age samples would need to be evaluated first, to ensure that they cover the likely stock distribution. If age-length keys can be generated for commercial data, then further work could explore the possibility of an age-based assessment model such as SAM. Concurrent developments in spatial length-based assessment methods (in Denmark and Scotland) could also be used to indicate stock trends in the absence of age estimates, and a variety of data-limited assessment methods could be explored as exploratory analyses.

No biological reference points are defined yet. The benchmark should explore if reference points can be defined or if the use of alternative indicators such as SSB proxies is possible.
The key first task of the benchmark will be to determine whether sufficient historical data exist to warrant a move towards a full analytic assessment, and to evaluate whether management of the stock on this basis would improve the efficacy of decisionmaking over the existing data-limited approach.

## Workplan

Compilation of catch data in InterCatch format (years to be confirmed): all national institutes; prior to the benchmark.

Compilation of survey data (IBTS and BTS): probably via DATRAS, although would also need to check that biological information for lemon sole has been uploaded. Scotland in collaboration with contributing institutes; prior to the benchmark.

Evaluation of survey indices: Scotland; prior to the benchmark.
Compilation of input data for age- and length-based assessment models: relevant counties; prior to the benchmark.

Exploratory assessment runs: Scotland; during the benchmark.

## Other working groups to be involved:

WGBEAM, IBTS-WG (age based indices, index uncertainty estimate, combination of IBTS and BTS indices).

| Stock | LEM-NSEA |  |
| :--- | :--- | :--- |
| Stock <br> coordinator | Name: Coby Needle | Email: needlec@marlab.ac.uk |
| Stock assessor | Name: Coby Needle | Email: needlec@marlab.ac.uk |
| Data contact | Name: Coby Needle | Email: needlec@marlab.ac.uk |

## Flounder in Subarea 4 and Divisions $3 a$

Data needed

- Catch (landings and discards) data back to 2002 (BEL, NDL, GER, DNK, SWE, NOR, UK ENG, UK, SCO, FRA).
- Effort data by single métiers as potential additional source of information to estimate the amount of discards where no sufficient information from sampling programs is available.
- Available IBTS, BTS and DYFS/DFS survey data on flounder.
- Available biological data from scientific surveys and sampling programs of the commercial fleets (length distributions, age distributions, spatial distributions, weight, maturity, sex ratios).


## Current assessment issues

Currently flounder is treated as a data limited species and the stock perception is derived from simple survey trends and catch data. For the current assessment method the IBTS-Q1 index (excluding round fish area 1 and 2; mature biomass index in kg ) is used as basis for the trend based analysis (DLS 3.2. method). The IBTS surveys appear to cover most of the stock distribution, but may not provide representative indices as the GOV gear is not optimized for flatfish sampling. The available beam-trawl surveys should provide a more representative index. Flounder is more distributed near coastal
areas, therefore also the inshore surveys (DYFS/DFS) should be evaluated as a possible data source for a representative survey index. Survey data needs to be evaluated to estimate a reasonable survey index for flounder. Further, it was often argued within the WGNSSK that the indices used for the DLS 3.2. method do not come along with uncertainty estimates and that these estimates should be calculated.
For commercial catches to be used in assessments, discards need to be quantified for flounder. Data are available to do this for 2012-2015, but more discard data sampled under the DCF should be available and should be uploaded into the InterCatch data portal by all relevant institutes.
Although flounder is treated as a data-limited stock, much data may exist which are not utilized today. The compilation of such data can possibly provide biological data for a number of years (including lengths, ages, maturity, weights and spatial location), and national catch-sampling programmes may be able to provide further valuable information. The compilation of these data will be one of the major tasks during the proposed benchmark.
No biological reference points are defined yet. The benchmark should explore if reference points can be defined or if the use of alternative indicators such as SSB proxies is possible.

## Workplan

Compilation of catch data in InterCatch format (years to be confirmed): all national institutes; prior to the benchmark.

Compilation of survey data (IBTS, BTS, DYFS/DFS): probably via DATRAS, although would also need to check that biological information for flounder has been uploaded: all national institutes.

Evaluation of survey indices; prior to the benchmark.
Compilation of input data for age- and/or length-based assessment models: relevant countries; prior to the benchmark.

Exploratory assessment runs: during the benchmark.

Other working groups to be involved:
WGBEAM, IBTS-WG (indices, index uncertainty estimate, combination of IBTS and BTS indices).

| Stock | fle-nsea |  |
| :--- | :--- | :--- |
| Stock coor- <br> dinator | Name: Holger Haslob | Email:holger.has- <br> lob@thuenen.de |
| Stock asses- <br> sor | Name: Holger Haslob | Email:holger.has- <br> lob@thuenen.de |
| Data contact | Name: Holger Haslob | Email:holger.has- <br> lob@thuenen.de |


| Issue | Problem/Aim | Work needed / possible direction of solution | Data needed to be able to do this: are these available / where should these come from? | External expertise needed at benchmark type of expertise / proposed names |
| :---: | :---: | :---: | :---: | :---: |
| New) data to be considered and/or quantified | Lemon sole have never been the subject of a full analytic assessment. A key role of the benchmark is therefore to determine whether data exist to enable an assessment of this kind. | See below. | See below. | See below. |
| Tuning series | Tuning series do not yet exist for lemon sole. | Age- or lengthbased tuning series should be generated on the basis of DATRAS data. | Data should be available in DATRAS, but national institutes should also be approached to determine if all relevant data have been uploaded. | Coby Needle (Sco), Liz Clarke (Sco), ICES DATRAS staff. |
| Discards | Discard estimates from 2013-2015 indicate average rates of around $30 \%$. Therefore, any catch-based assessment will need to account for discards. | Check availability of discard data from commercial sampling programmes and upload data to InterCatch for years prior to 2013. | Discard information from national sampling programmes. <br> All relevant institutes (BEL, DNK, NDL, GER, ENG, SCO, FRA, SWE, NOR). | Coby Needle (Sco), Liz Clarke (Sco). |
| Biological <br> Parameters | To collate and compile available data on weight, length, maturity, age, sex and spatial distribution. | Standard approaches currently applied to stocks such as haddock and plaice could be applied to collate these data. | Much of the required information can be obtained from DATRAS, but national institutes also need to be approached about the availability of relevant (and unsubmitted) data from survey and catchsampling programmes. | Coby Needle (Sco), Liz Clarke (Sco), Rasmus Nielsen (Den). |


| Issue | Problem/Aim | Work needed / possible direction of solution | Data needed to be able to do this: are these available / where should these come from? | External expertise needed at benchmark type of expertise / proposed names |
| :---: | :---: | :---: | :---: | :---: |
| Assessment method | Lemon sole are not currently assessed using an analytic method. | The applicability and utility of a range of candidate models to lemon sole needs to be evaluated. | The models to use depends on the data available (see previous row). | Coby Needle (Sco), Anders Nielsen (Den), Tanja Buch (Den), Colin Millar (ICES) |
| Biological <br> Reference <br> Points | No biological reference points exist for lemon sole. | The approach used to determine reference points will depend on the data available and the assessment methods used. | See above. | Coby Needle (Sco), Tanja Miethe (Sco), Alex Kempf (Ger). |

## PART B

Stock Data Problems Relevant to Data Collection -WGNSSK

| Stock | Data Problem | How to be ADDRESSED in | By who |
| :---: | :---: | :---: | :---: |
| Stock name | Data problem <br> identification | Description of data problem |  |
|  |  |  | Who should <br> and recommend solution <br> recommended <br> solution and <br> who should be |
| notified on this |  |  |  |
| data issue. |  |  |  |


| Ple-nsea, sol-nsea | An increasing number of beam trawlers (in the Dutch fleet) are using 'Pulse trawl' gear. There is no recognised gear code for this gear and catches etc. are still registered as TBB, grouping them with the traditional twin beam trawl fleet. | It is felt that this gear is likely to have different selectivity (for discards and landings) as well as different catch per unit effort as the traditional beam trawl gears. This has implication for the assessment of sole and plaice. In the first case, for the raising of discards and landings data. In the second case for the determination of the CPUE index used in the sole assessment. It is necessary to create a separate gear code / gear type category for pulse trawls. This would allow for improved raising of data and prevent a discontinuity in the CPUE index for sole. | RCM-NS\&EA, RBD-SG |
| :---: | :---: | :---: | :---: |
| Saithe in <br> Subarea IV, <br> VI and <br> Division <br> IIIa | No accoustic survey index for older yearclasses, assessment heavily dependent on commercial CPUE | The NORACU can no longer be used in the assessment because of errors in sampling design and inconsistencies in the time series. Establish an acoustic survey in Q1 or Q3 to get fishery independent information on older age groups . | ACOM <br> (Norway); <br> ACOM <br> (Germany); <br> ACOM (France), <br> ACOM <br> (Denmark); <br> ACOM <br> (Scotland) |
| Saithe in <br> Subarea IV, <br> VI and <br> Division <br> IIIa | No recruitment index time series | The number of recruits is difficult to determine before they have been targeted by the fishery. Establish a recruitment survey . | ACOM (Norway) |
| Saithe in <br> Subarea IV, <br> VI and <br> Division <br> IIIa | Age sampling from commercial fleets | Possible cluster sampling due to few vessels in the reference fleet (Norway), needs review / redesign | ACOM <br> (Norway); <br> PGDATA |


| Stock | Data Problem | How to be Addressed in | By wно |
| :---: | :---: | :---: | :---: |
| Turbot in IIII, | Small turbot stocks cannot be easily assessed because of potentially large migrations in and out the large areas IV and the Baltic. | Most knowledge about stocks connectivity is based on old and limited tagging experiments. New tagging studies would be necessary to improve the understanding of migratory patterns | SIMWG; ACOM <br> (Denmark, <br> Sweden) |
| Nep 32 | Deficient <br> Norwegian catch sampling | The coast guard sampling of Norwegian and Danish commercial catches is satisfactory in some years, but not in others. The main problems with these data are that catches are often measured by total length (whole cm) and sample weight is missing. As total length data have lower resolution compared with carapace length data, the two cannot be combined without losing accuracy. The coast guard is aware of these problems and strives to improve the data | ACOM <br> (Norway) |
| $\text { Nep } 32 \&$ IIIa | Scarce <br> Norwegian log book data | The Norwegian logbook system was changed in 2011 with the introduction of electronic logbooks compulsory for all vessels $\geq 15 \mathrm{~m}$. In 2013 compulsory electronic logbooks for vessels $\geq 12 \mathrm{~m}$ were introduced in FU 3. As a large portion of the Norwegian fleet landing Nephrops in FU 3 and 32 consists of vessels $<12 \mathrm{~m} /<15 \mathrm{~m}$, the logbook data will continue to be limited. <br> A growing part of the Norwegian Nephrops landings come from the trap fishery, but this part of the fleet is not required to fill in logbooks, probably because of the small size of the vessels. Logbooks from traps would provide data from the eastern (less exploited) part of FU 32. <br> Log books should be introduced for vessels $<15 \mathrm{~m}$, including trap fishers. | ACOM <br> (Norway) |
| Pollack in Subarea IV and Division IIIa | General lack of biological data needed for better understanding of growth and maturity. | In routine surveys, such as the quarter 1 and quarter 3 IBTS in Subarea IV and Division IIII, apart from reporting catches at length, no biological data are collected for this species. In order to understand better their growth and maturity WGNEW recommended that otoliths and maturity information should be collected during these surveys for a few years. WGNSSK also recommends that biological data from commercial catches should be processed. | IBTSWG; RCM- <br> NS\&EA |


| Stock | Data Problem | How to be Addressed in | BY wнo |
| :---: | :---: | :---: | :---: |
| Whiting in Division IV and IIIa | General lack of stock identity and area specific age readings | Studies on whiting stock identity and connectivity in western Baltic, Division IIIa and Division IV should be encouraged In the routine surveys, IBTS quarter 1 and quarter 3 in Division IIIa, apart from reporting catches at length, no biological data are collected for this species. In order to understand better their growth and maturity it is recommendable that otoliths and maturity (also in area IV) information should be collected during surveys. | National research services and IBTSWG |
| Cod in subdivision IIIaW, subarea IV, and division VIId | Perceived catchability problems in IBTS Q1 and Q3 indices, | Appropriate standardisation of IBTS Q1 and Q3 surveys was carried out during WKNSEA 2015. Inconsistencies were found between q1 and q3 in the Skagerrak area. However, so far only one vessel is fishing in the Skagerrak making it impossible to differentiate vessel, gear and crew effects from real changes in abundance. It is recommended that also in the Skagerrak two vessels fish in each ICES rectangle. This is the standard in all other areas covered by the IBTS. | IBTS-WG, <br> ACOM <br> (Danmark, Sweden, Germany, Norway). |
| Nephrops <br> FU 33 | Not enough discard information available to give catch advice | The sampling in this FU is insufficient. Samples are needed from the main fleets fishing in this FU. | ACOM <br> (Denmark, <br> Netherlands, <br> Belgium, <br> Germany) |
| Turbot in IV | Biological information is only available from the Netherlands. This is a serious concern leading to a potentially biased assessment | Age information is needed also from other countries. So far age distributions are mainly available from the Dutch BT2 fishery. However, these samples may not be representative for other fisheries and countries (e.g., gill net fishery, otter trawl fisheries). All available information needs to be uploaded to Intercatch as far back in time as possible. Future sampling effort needs to ensure a proper sampling coverage over the main fleets and countries. | ACOM <br> (Denmark, UK, <br> Germany, <br> Belgium) |


| Stock | Data Problem | How to be addressed in | ВY wнo |
| :---: | :---: | :---: | :---: |
| Sole-eche | The UK YFS stopped in 2006 and the French Young Fish survey as conducted now is probably not providing the correct recruitment estimates as it only covers part of VIId | The UK component of the YFS index is not available since 2007, resulting in the unavailability of the combined YFS-index. This combined index has been estimating the incoming year class strength very consistently, hereby providing reliable estimates to the forecasts. Although results of using the YFS indices separately (FR-YFS for 1987-present and UK-YFS for 1987-2006) did not show apparent changes in retrospective patterns, it was noted that the lack of information from the UK YFS will affect the quality of the recruitment estimates and therefore the forecast. In RCT3 analysis the FR-YFS gets hardly any weight and the gemoetric mean has to be used instead. Possible solutions could be that either the UK YFS is conducted again in future years or the French Young Fish survey can be extended to include at least some of the sampling points from the former UK Young Fish survey. | ACOM <br> (UK,France) |
| Nep 5 | Incomplete catch sampling | Only Dutch catches are sampled, and discard data were only available for 2015. Length distributions and sex ratios are poorly defined due to limited sampling. Acknowledging that this is a difficult fishery to effectively sample, electronic capture of atsea data could be developed. | ACOM (UK, <br> Netherlands, <br> Germany, <br> Belgium) |

## Annex 07 WGNSSK data calls

No information provided for 2016.

## Annex 08 Audit Reports

No information provided for 2016.

## Annex 09 Working documents

## Alternate SAM models

## DATRAS standard Q3 index

The standard DATRAS Q3 indices (DATRAS indies), ages 3-8, 1992-2015 were used instead of the GAM generated indices in the SAM assessment model. Figure 1 shows the indices and the internal consistencies. The standard indices do not include the Skagerrak/Kattegat, but do include the southern North Sea (where saithe are no found). The truncated GAM-derived Q3 index (no Skagerrak/Kattegat or southern NS were compared with the DATRAS estimates for the expanded age range (Figure 2). The GAM and standard DATRAS indices are (generally) very similar. However, ages 4 and 5 , especially in the last 2 years, are over-estimated by the GAM (especially age 5).

Results of the SAM assessment are in Figure 3. Estimated SSB using the DATRAS indices closely mirrors estimates from the cpue-only model until around 2010, unlike the model with Q3 indices estimated with the GAM model. The DATRAS model shows slightly lower SSB than the Q3 GAM indices in 2015. Retrospective patterns show that SSB has been consistently underestimated, while fishing mortality has been mostly over-estimated. The retrospective patterns in $\mathrm{F}_{4-7}$ are not as poor as the model with the Q3 GAM indices. Residual plots are in Figure 5. Estimated catchabilities were very low for the DATRAS model, compared to the indices estimated with the GAM model.

## GAM Q3 index but without 2015

Results of the SAM assessment are in Figure 6. Omitting the 2015 Q3 data resulted in SSB2015 estimates lying between those estimated by the cpue-only model and the GAMestimated Q3 (with 2015) model. Retrospective patterns are in Figure 7; retrospectives are worse than the model including 2015 data (Figure 8). Residual patterns are in Figure 9.

## GAM Q3 index but with stock weights=catch weights for ages 7-10+

Not finished - bounds for stock weights
Results of the SAM assessment are in Figure 10. Replacing stock weights with catch weights for ages 7-10+ (where stock weights were greater than catch weights) made a large difference in the SAM output. While SSB still increases in the last two years of the series, SSB is lower for this model until 2014 than all other models. Retrospective patterns are in Figure 11. Residual patterns are in Figure 12.

## DATRAS Q3 index but with stock weights=catch weights for all ages

Results of the SAM assessment are in Figure 13; this is the model recommended as the final model based on the external review in early June. Replacing stock weights with catch weights for all ages had the effect of increasing SSB in comparison with the model where stock weights were replaced for ages 7-10+. This is because for ages 3-6, catch weights are higher than stock weights (Figure 14); these are the fish the make up the dominant part of the catch for the targeted trawl fisheries. Retrospective patterns are in Figure 15 and residuals are in Figure 16.


Figure 1. Standard DATRAS indices for Q3, 1992-2015, ages 3-8 and internal consistencies.


Figure 2. Standard DATRAS indices for Q3, 1992-2015, ages 3-8.


Figure 3. Trends in SSB, F4-7, and recruitment for the 4 models. Blue line: Q1 + GAM-estimated Q3 + cpue index model; green line: cpue-only model; orange line: combined cpue + GAM-estimated Q3 (truncated to exclude Skagerrak/Kattegat and southern NS); black line: DATRAS Q3 + cpue model; orange/tan shaded region: $95 \%$ confidence interval for the DATRAS Q3 + cpue model; solid grey line (grey dashed confidence intervals) are the previously saved base model (unknown at this point).


Figure 4. Eight year retrospective pattern in SSB, $\mathrm{F}_{4-7}$, and recruitment. Model is combined cpue + DATRAS Q3 and includes the discard revisions.


Figure 5. Residual patterns for the combined cpue + DATRAS Q3 assessment model. (left) Before correlation taken into account between ages, within years in the Q3 index; (right) after accounting for the correlation.


Figure 6. Trends in SSB, F4-7, and recruitment for the 5 models. Blue line: Q1 + GAM-estimated Q3 + cpue index model; green line: cpue-only model; orange line: combined cpue + GAM-estimated Q3 (truncated to exclude Skagerrak/Kattegat and southern NS); purple line: DATRAS Q3 + cpue model; black line (orange/tan shaded region: $95 \%$ confidence interval): GAM-estimated Q3 indices without 2015 + cpue model; solid grey line (grey dashed confidence intervals) are the previously saved base model (unknown at this point).


Figure 7. Eight year retrospective pattern in SSB, F4-7, and recruitment. Model is combined cpue + GAM-estimated Q3 (without 2015, excludes Skagerrak/Kattegat and southern North Sea) and includes the discard revisions.


Figure 8. Eight year retrospective pattern in SSB, $\mathrm{F}_{4-7}$, and recruitment. Model is combined cpue + GAM-estimated Q3 (excludes Skagerrak/Kattegat and southern North Sea) and includes the discard revisions.


Figure 9. Residual patterns for the combined cpue + GAM-estimated Q3 assessment model (no 2015). (left) Before correlation taken into account between ages, within years in the Q3 index; (right) after accounting for the correlation.


Figure 10. Trends in SSB, F4.7, and recruitment for the 5 models. Blue line: Q1 + GAM-estimated Q3 + cpue index model; green line: cpue-only model; orange line: combined cpue + GAM-estimated Q3 (truncated to exclude Skagerrak/Kattegat and southern NS); purple line: DATRAS Q3 + combined cpue; black line (orange/tan shaded region: $95 \%$ confidence interval): GAM-estimated Q3 indices + cpue model + stock weights=catch weights for ages 7-10+; solid grey line (grey dashed confidence intervals) are the previously saved base model (unknown at this point).


Figure 11. Eight year retrospective pattern in SSB, $\mathrm{F}_{4-7}$, and recruitment. Model is combined cpue + GAM-estimated Q3 (excludes Skagerrak/Kattegat and southern North Sea) + stock weights=catch weights for ages 7-10+ (includes discard revisions).


Figure 12. Residual patterns for the combined cpue + GAM-estimated Q3 + stock weights=catch weights for ages 7-10+ assessment model. (left) Before correlation taken into account between ages, within years in the Q3 index; (right) after accounting for the correlation.


Figure 13. Trends in SSB, $\mathrm{F}_{4-7}$, and recruitment for various models. Blue line: (benchmark model) Q1 + GAM-estimated Q3 + combined cpue index model; green line: combined cpue-only model; orange line: combined cpue + GAM-estimated Q3 (truncated to exclude Skagerrak/Kattegat and southern NS); purple line: GAM-estimated Q3 indices + combined cpue model + stock weights=catch weights for ages $7-10+$; black line (orange/tan shaded region: $95 \%$ confidence interval): DATRAS Q3 indices + combined cpue, stock weights=catch weights for all ages.


Figure 14. Stock weights (dashed lines) and catch weights (solid lines) for ages 1-10+. The left panel shows age 3 (black lines) to age 6 (light blue lines), while ages 7-10+ are in the right panel. This figure differs from the benchmark working document due to re-raising (InterCatch bug and changing of raising procedure for Norwegian discards).


Figure 15. Eight year retrospective pattern in SSB, $\mathrm{F}_{47}$, and recruitment. Model is combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat) + stock weights=catch weights for all ages (includes discard revisions).

|  | Total catches |
| :---: | :---: |
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| $\infty$ |  |
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Figure 16. Residual patterns for the combined cpue + DATRAS Q3 (excludes Skagerrak/Kattegat) (stock weights=catch weights for all ages) assessment model. (left) Before correlation taken into account between ages, within years in the Q3 index; (right) after accounting for the correlation.

# CGFS : Change of vessel from 2015 onwards and consequences on survey design and stock indices 

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## Introduction

The Channel Ground Fish Survey (CGFS) has been conducted in the eastern English Channel yearly in October since 1988 with a systematic fixed sampling program. The CGFS was realized using a high opening (GOV) bottom trawl ( 20 mm meshsize coden) and 30 minutes trawls using the same RV Gwen Drez since 1988.

The RV Gwen Drez was decommissioned in 2015 but given the international importance of the CGFS it was decided to continue the time series using the RV Thalassa. In order to allow for a continuation of the time series an intercalibration was realized in 2014 by conducting paired tows, simultaneously with both vessels (see appendix of the WGIBTS 2015 report for description of the intercalibration results).

## Adaptation of the sampling design

## Rationale

Based on the characteristics of the new RV Thalassa (bigger draught), and the vessel time availability at this period of the year, three scenarios of reduction of the trawling stations set have been tested. For each scenario, a selection of hauls was made among the 89 hauls of original sampling scheme of the survey (Fig. 1) based on different criteria. The relevance of these subsets of hauls was assessed by resampling on the historical time series and computing the associated abundance indexes per age for plaice (Pleuronectes platessa) and red mullet (Mullus surmuletus) which are assessed using this survey as tuning fleet in the ICES WGNSSK (Working Group on North Sea, Skagerrak and Kattegat).


Figure 1 : Hauls of the original CGFS sampling scheme

## Selected scenario

After a trial and error experiment on the haul selection, the selected sampling scheme include the areas easily fishable by the RV Thalassa ( 69 hauls that are always more than 15 meters deep) and some of the shallower hauls (limited to the ones which the average bathymetry over the time series is over 15 meters: 5 hauls). This selection allowed for covering 74 hauls (i.e. $83 \%$ of the initial sampling scheme, Ann.1). It excludes the hauls outside VIId which were not used.

To test the relevance of the selected hauls, the internal consistency of the indices was tested for two species (plaice and red mullet).

## Red mullet



Figure 2 : Internal consistency ; 2a : New index based on the subset of hauls, 2 b : reference index
Correlation coefficients appeared to be higher with this selection of hauls than with the original sampling scheme, improving the internal consistency of the index for red mullet.

## Plaice



Figure 3 : Internal consistency; 3a: New index based on the subset of hauls, 3b: reference index

The internal consistency of the original index was not completely satisfying. The new sampling scheme resulting from a subset of the original stations do not deteriorate this internal consistency further. The correlation coefficients are of the same order of magnitude for ages $2 / 3,3 / 4$ and slightly increased for ages $4 / 5$.


Figure 3 : New spatial coverage of the Channel Ground Fish Survey

## From CPUE in number per hour fished to CPUE in number per $\mathrm{km}^{2}$

The original index provided was computed in number of fish per hour fished. In a first step an index was computed per ICES square (the stratum in this survey) and then elevated to the whole Eastern Channel to compute a number of fish per age class and hour fished.

As the surface trawled differed between the two RV (difference in trawling speed and width of the gear used $\left(0.029 \mathrm{~km}^{2}\right.$ on average for the RV Gwen Drez over the period 2008-2014 against $0.052 \mathrm{~km}^{2}$ for the RV Thalassa (average of the hauls realized in 2015)) a density index (number of fish per $\mathrm{km}^{2}$ ) was also tried in order to create a consistent index over the whole time series. This is in line with the current effort led by the IBTSWG to produce trawled surface and density indices for all the expert groups for 2017.

The index is then computed using the formula:

With :

[^11]$$
\bar{N}=\frac{\sum_{s} A_{s} \cdot \overline{N_{s}} \quad A_{s} \text { Surface of the strata s, in } \mathrm{km}^{2} \text {, } \sum_{s} A_{s}}{\text {. }}
$$

As the vertical opening of the gear used by the RV Thalassa was higher than the previous one, and in order to take into account any vessel effect on catchability, the CPUE were compared for all the species caught. Differences in CPUEs between the new and the old survey setting were found for 9 species (mostly pelagic species). In the case of plaice and red mullet, CPUEs were not significantly different, so no conversion factor was applied to these two species.

## Differences in indices provided in 2015 and 2016

During the process of automatizing the computation of the index, some errors were found in the surface of some strata and ALK used for some species. These errors where corrected and the new indices (expressed in number of fish per $\mathrm{km}^{2}$ instead of number of fish per hour fished) take these corrections into account.

In order to compare the "old" and "new" CGFS indices for plaice and red mullet they were first plotted against each other to get a visual comparison of the index values at age and assess the possible differences and inconsistencies. The correlations between indices at age time series were then computed to check for consistency between these two indices. The last step was to check the internal consistency to assess the impact of the new calculation on the indices.

## Comparison Old/New index for plaice

Index at age


Figure 4 : CGFS old (blue) and new (black) standardized index at age

The main trends of the CGFS index at age remain very similar. The main differences are:

- for age 1 in 1997 where the peak observed is no longer observed with the new calculations;
- a new peak in the age 1 in 2011 which is in line with what was observed by the UK-BTS survey that year;
- the main differences are observed for age 6 , where the two indices seem to be inconsistent ;
- for age 0 in 2000 and age 1 in 2011, two new peaks appeared with the new calculation.

In the assessment only the ages 1 to 6 are used.

## Correlations between the two different indices at age time series



Figure 5 : Correlation between indices at age for the old and new indices
The correlations for ages $0,2,3,4$ and 5 are high, reflecting the coherence seen when plotting the old and new surveys against each other. Correlations for ages 1 and 6 are weaker, also reflecting the differences for some years for age 1 and a poor consistency between new and old indices for age 6 .

## Internal consistency

FR GFS_new


Lower right panels show the Coefficient of Determination $\left(r^{2}\right)$

FR GFS_old

$\log _{10}$ (Index Value)
Lower right panels show the Coefficient of Determination $\left(r^{2}\right)$

Figure 6: Internal consistency for new and old indices
The internal consistency is globally improved. Correlation coefficients are increased for ages $1 / 2$ and $4 / 5$ while they do not vary much for ages $2 / 3$ and $3 / 4$.

## Comparison Old/New index for mullet

Index at age


Figure 7: CGFS old (black) and new (blue) standardized index at age
The main trends of the CGFS index at age are remaining very similar. The main differences are for age 2 in 2008 where the peak observed in the new calculations is higher than the one from the old index. For Age 4 in 2004 and 2005, indices seem to be inconsistent with a decrease between 2005/2006 whereas the index increased with the old index.

## Correlations between the two different indices at age time series



## Figure 8

## Correlation between indices at age for the old and new indices

The correlations for all ages are high but with very few data points after age 4.
Internal consistency


Figure 9: Internal consistency for new and old indices

The main patterns are maintained from the old to the new index. The higher correlation is between age 2 and 3 but increased with the new index.

## Annex 1: Hauls kept in the new survey

| Trait_Selection_CgFs_Thalassa_S3 |  |  |  |
| :---: | :---: | :---: | :---: |
| sta_Recodage | Latitude | Longitude | Trait2014 |
| 4M1 | 50.195 | 1.171667 | 62 |
| 4M2 | 50.01667 | 1.083333 | 67 |
| 5L1 | 50.385 | 0.7916667 | 84 |
| 5M1 | 50.305 | 1.171667 | 61 |
| 6 J 1 | 50.54 | 0.3533333 |  |
| 6K1 | 50.56167 | 0.7283334 | 57 |
| 6L1 | 50.53667 | 0.89 | 59 |
| 2D2 | 49.59 | -1.118333 | 32 |
| 2E1 | 49.58333 | -0.9666666 | 34 |
| 3D1 | 49.82833 | -1.135 | 1 |
| 3E1 | 49.99833 | -0.7683333 | 2 |
| 3F1 | 49.96 | -0.6216667 | 17 |
| 3H1 | 49.90333 | -0.1216667 | 36 |
| 3 I 1 | 49.84333 | 0.1633333 | 71 |
| 3 J 1 | 49.87833 | 0.4333333 | 70 |
| 3K1 | 49.895 | 0.5116667 | 69 |
| 3L1 | 49.98167 | 0.825 | 68 |
| 4C1 | 50.04 | -1.275 | 96 |
| 4D1 | 50.09833 | -1.265 | 4 |
| 4E1 | 50.02833 | -0.905 | 3 |
| 4F1 | 50.075 | -0.6183333 | 15 |
| 4G1 | 50.08833 | -0.4566667 | 14 |
| 4H1 | 50.245 | -0.05 | 10 |
| 411 | 50.01833 | 0.175 | 37 |
| 4J1 | 50.09333 | 0.3266667 | 83 |
| 4K1 | 50.11333 | 0.6016667 | 41 |
| 5D1 | 50.415 | -1.166667 | 76 |
| 5E1 | 50.47667 | -0.9066667 | 74 |
| 5F1 | 50.44333 | -0.5966667 | 72 |
| 5H1 | 50.34667 | -0.1583333 | 12 |
| 5 I 1 | 50.355 | 0.005 | 11 |
| 5J1 | 50.30167 | 0.4133333 | 86 |
| 5K1 | 50.35833 | 0.6366667 | 85 |


| Trait_Selection_CgFs_Thalassa_S3 |  |  |  |
| :---: | :---: | :---: | :---: |
| sta_Recodage | Latitude | Longitude | Trait2014 |
| 6E1 | 50.52333 | -0.8833333 | 75 |
| 6F1 | 50.525 | -0.71 | 73 |
| 6G1 | 50.57333 | -0.4433333 | 80 |
| 1D1 | 49.42333 | -1.058333 | 29 |
| 2D1 | 49.51167 | -1.223333 | 97 |
| 2 I | 49.64 | 8.166666E-02 | 88 |
| 2 I 2 | 49.60167 | $8.333334 \mathrm{E}-02$ | 89 |
| 1E1 | 49.42333 | -0.985 | 28 |
| 1E2 | 49.45 | -0.9233333 | 27 |
| 1F1 | 49.46167 | -0.675 | 26 |
| 1F2 | 49.41667 | -0.5533333 | 25 |
| 1G1 | 49.45833 | -0.43 | 24 |
| 1G2 | 49.47167 | -0.325 | 23 |
| 2F1 | 49.66167 | -0.6433333 | 35 |
| 2G1 | 49.55667 | -0.3316667 | 20 |
| 2H1 | 49.65333 | -0.145 | 19 |
| 3G1 | 49.84 | -0.2533333 | 18 |
| 1H1 | 49.46333 | -0.1433333 | 22 |
| 1H2 | 49.35833 | -0.1766667 | 90 |
| 7 O 2 | 50.79 | 1.558333 | 49 |
| 7G1 | 50.76 | -0.2833333 | 79 |
| 7K1 | 50.79167 | 0.5333334 | 54 |
| 7L1 | 50.87667 | 0.8366666 | 53 |
| 7L2 | 50.78167 | 0.84 | 56 |
| 7M1 | 50.97 | 1.085 | 51 |
| 6 O 1 | 50.655 | 1.541667 | 42 |
| 7 O | 50.91333 | 1.61 | 48 |
| 7H1 | 50.755 | -0.1216667 | 78 |
| 7N1 | 50.86666 | 1.346667 | 50 |
| 6H1 | 50.56 | -0.1266667 | 81 |
| 6 I1 | 50.635 | 7.666667E-02 |  |
| 3M1 | 50.00834 | 1.218333 | 65 |
| 4N1 | 50.2 | 1.39 | 63 |
| 4N2 | 50.09 | 1.37 | 64 |
| 5N1 | 50.47167 | 1.438333 | 46 |
| 5N2 | 50.41833 | 1.345 | 47 |
| 5 O 1 | 50.44833 | 1.526667 |  |
| 6M1 | 50.66 | 1.005 | 58 |

## Trait_Selection_Cgfs_Thalassa_S3

| sta_Recodage | Latitude | Longitude | Trait2014 |
| :--- | :--- | :--- | :--- |
| 6 N 1 | 50.57667 | 1.43 | 44 |
| 6 O 2 | 50.56333 | 1.51 | 43 |
| 4L1 | 50.15667 | 0.9766667 | 60 |

# Working Document to the ICES Working Group on the Assessment of Demersal Stocks in the North Sea and Skagerrak (WGNSSK), April 2016 

Not to be cited without prior reference to the author

Plaice (Pleuronectes platessa) and sole (Solea solea) in the UK Beam trawl survey in the Eastern English Channel (7d)

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#### Abstract

The present document describes the calculation of the plaice and sole survey indices in the Q3 UK beam trawl survey in the ICES division 7d. Further investigations were made in relation to the survey data quality, selection of prime stations, age-at-length keys and ultimately the indices calculations. Results for revisited index and previous index are shown in the present document to facilitate a better comparison and further inform on temporal differences to the year class abundance.


## Survey indices

The present document describes the calculation of the plaice and sole indices in the UK Q3 beam trawl survey in the Eastern English Channel and Southern North Sea. The annual procedure is currently done automatically using the Cefas Fishing Survey System (FSS) and R software, and provides the index for the time-series since 1989 and 1996 (7d and 4c, respectively). Prior to 2016, survey indices were calculated using Cefas FSS, SAS code and Microsoft Excel® outputs. Whilst re-writing the code from SAS to $R$, discrepancies were found in the selection of valid primes used in the production of age-length keys (ALKs) and length-distributions (LDs), with survey data within Cefas FSS revisited and corrected accordingly, where possible. It should also be noted, current survey biological sampling targets (otoliths) for both species are set by sector (7d UK Inshore, 7d France Inshore and 4c North Sea), though previous indices had calculated ALKs by ICES rectangles.

Therefore, this document refers to an update of index calculations so that they are consistent with current survey data collection protocols. Data prior to 2005 presented for the 7 d area should be viewed and used with some caution, since these data were not revisited and reviewed in terms of their quality, and historical data collection procedures may differ from the current one.

New results for survey area 4c were also provided to the 2016 ICES WGs, and although are not discussed in the present document, should be viewed only as provisional, because further investigations are required on the survey data and historical prime selection when current primes were not fished.
A total of 75 primes ( 39 in the UK and 36 in the FR sector), were selected from 19892015, with a few currently not fished, though historically relevant (Figure 1). Primes
used for the length-distributions (LD) and fishing effort are within the UK sector, 22-$27,42-45,47,50-67,73-75$ and 94 , with $1,4,6-12,16-21,29,35-40,68-72,76-77$ and 95 within the FR sector. Primes $2,3,5,14,15,41$ and 46 are included only in the age-length key (ALK) as they have not been fished in recent years, though historically were part of the survey primary grid. Similarly, only included in the ALKs calculations are primes 200, 201, 202 and 203 within the UK sector. These are set as additional and no longer fished since 2014, though historically, otoliths have been collected as part of the survey target. It should be noted that the prevalence of static gear around prime 49, currently on the main survey grid, has prevented the tow to be fished successfully in recent years. Therefore, data collected for the latter prime has been excluded from LDs and fishing effort, and only used as part of the ALK for the UK stratum.

R code procedures include an initial data retrieval from Cefas FSS, where data are electronically stored, for the relevant prime stations where fishing operations were considered valid. Numbers at length for each fishing station are standardized to 30 minute tows, with the raising factor dependent on the actual tow duration. The total number across stations within an ICES rectangle results in the LD for ICES rectangle within sector (UK and FR). The ALK derived from the biological sampling at sea (otolith collection) is produced separately by UK and FR sector and raised to the appropriate LDs, resulting in an age-length composition for each ICES rectangle-sector-sex combination. The ALKs and LDs for plaice are calculated by sex for all years, and for sole calculated by sex when measured and biologically sampled by sex (1993-2009), and combined when measured and biologically sampled unsexed (1989-1992, 2010-onwards). The LDs used are only from valid stations; meanwhile ALKs use all stations within the chosen primes, even when considered additional or invalid tows to accommodate the occasional biological sampling occurrence. The total numbers across lengths by age create the age composition (AC) for each ICES rectangle-sector-sex combination, with the sum as the AC for the survey year. These are divided by the total number of valid primes fished across UK and FR sectors, which may differ from the number of primes with plaice and/or sole catches. The results are further raised and multiplied by four to give the final index equivalent to one-hour tows with an 8-metre beam trawl (the factor four is because stations are standardised to 30 minute tows and conducted with a 4 metre beam trawl).

Furthermore, the R code is designed to reallocate, where possible, miss-matches where a fish at a given length has no associated record in the ALK, with code reallocating numbers at length (LD) up to a maximum of $\pm 2 \mathrm{~cm}$ of the initial length. If there are age records either above or below the initial length group, the fish are reallocated to those respective length groups within the LD. However, if age records are found in both lengths above and below the initial length group, the fish are split between those two lengths groups, using the ratio of each value divide by the sum of both length groups. If code is unable to reallocate fish, data are not used for further index calculations.

## Results

The revised survey indices for plaice and sole in 7d area are presented on Table 1 and 3. Previous index provided to the WG is presented on Table 2 and 4. A comparison between the two indices is presented on Figure 2 and 3 so as to better inform through visualization if there are any substantial temporal changes on year class abundance for fish aged one to six (ages currently used in the assessment).

Overall long-term trends for plaice are similar between the two indices for 1-year to 6year class (Figure 2). Meanwhile, for sole, although the main increases and declines are
being picked up by the two indices, there may be few discrepancies with historical data (e.g. 1999) (Figure 3).

## References

R Core Team (2014). R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL http://www.R-project.org/

Table 1 - Revised index for plaice in the UK-7D BTS (1989-2015)

| Age/Year | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 4.39 | 1.30 | 0.00 | 0.00 | 0.00 | 0.20 | 0.00 | 24.14 | 0.98 | 43.19 | 1.38 | 1.59 | 2.73 | 1.31 |
| 1 | 3.79 | 9.24 | 16.80 | 22.37 | 4.59 | 9.35 | 14.48 | 22.09 | 48.17 | 30.59 | 12.82 | 19.53 | 27.90 | 37.86 |
| 2 | 15.84 | 9.39 | 14.53 | 21.31 | 20.18 | 8.54 | 6.24 | 17.26 | 28.55 | 37.93 | 10.67 | 30.19 | 20.27 | 25.86 |
| 3 | 28.93 | 11.13 | 11.47 | 6.60 | 7.99 | 10.07 | 3.80 | 1.73 | 10.97 | 12.06 | 28.77 | 18.75 | 14.12 | 12.51 |
| 4 | 31.66 | 11.73 | 8.68 | 6.64 | 2.79 | 5.95 | 5.68 | 1.03 | 1.25 | 4.98 | 4.62 | 20.47 | 9.82 | 5.46 |
| 5 | 4.00 | 12.59 | 8.64 | 7.17 | 2.87 | 1.98 | 2.22 | 2.00 | 1.57 | 0.63 | 1.61 | 4.99 | 14.84 | 2.62 |
| 6 | 1.72 | 1.53 | 4.60 | 5.41 | 2.38 | 0.61 | 0.75 | 1.29 | 0.51 | 0.60 | 0.31 | 1.27 | 2.74 | 5.28 |
| 7 | 1.65 | 0.96 | 1.83 | 3.20 | 3.05 | 0.97 | 0.75 | 0.57 | 0.56 | 0.65 | 0.19 | 0.73 | 0.78 | 0.98 |
| 8 | 0.63 | 1.23 | 1.08 | 0.54 | 3.42 | 1.73 | 1.48 | 0.38 | 0.36 | 0.32 | 0.26 | 0.38 | 0.45 | 0.20 |
| 9 | 0.31 | 1.02 | 0.11 | 0.28 | 0.62 | 1.78 | 1.17 | 0.66 | 0.20 | 0.30 | 0.13 | 0.44 | 0.32 | 0.17 |
| $10+$ | 1.75 | 0.63 | 1.14 | 0.79 | 0.65 | 0.80 | 1.36 | 4.13 | 1.84 | 2.03 | 1.01 | 2.04 | 1.79 | 0.90 |
| Total (ages 1-10+) | 90.27 | 59.44 | 68.87 | 74.30 | 48.53 | 41.77 | 37.93 | 51.12 | 93.98 | 90.10 | 60.39 | 98.79 | 93.04 | 91.83 |
| Age/Year | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |  |
| 0 | 3.20 | 15.97 | 0.34 | 5.58 | 0.23 | 0.13 | 8.76 | 1.36 | 12.30 | 0.00 | 0.22 | 0.52 | 0.00 |  |
| 1 | 10.62 | 52.93 | 15.62 | 30.06 | 53.11 | 39.58 | 77.73 | 64.24 | 115.07 | 24.69 | 32.26 | 145.33 | 37.99 |  |
| 2 | 39.70 | 22.48 | 36.18 | 28.85 | 28.90 | 40.58 | 39.53 | 64.70 | 112.22 | 81.10 | 61.02 | 156.47 | 178.70 |  |
| 3 | 9.81 | 20.72 | 12.80 | 16.80 | 12.17 | 10.51 | 20.92 | 17.74 | 39.55 | 55.98 | 88.19 | 50.67 | 63.19 |  |
| 4 | 4.42 | 4.75 | 10.04 | 5.94 | 6.21 | 4.29 | 5.87 | 9.15 | 10.28 | 18.65 | 45.04 | 62.13 | 30.15 |  |
| 5 | 2.28 | 1.15 | 3.19 | 4.27 | 3.17 | 3.84 | 3.23 | 3.12 | 7.00 | 4.24 | 10.24 | 26.75 | 33.42 |  |
| 6 | 1.14 | 0.26 | 1.07 | 1.31 | 2.90 | 1.80 | 2.27 | 1.72 | 2.85 | 3.30 | 3.41 | 8.95 | 15.69 |  |
| 7 | 2.67 | 0.84 | 0.64 | 1.08 | 0.82 | 0.90 | 0.77 | 1.27 | 1.09 | 1.06 | 1.13 | 1.96 | 3.30 |  |
| 8 | 0.81 | 1.27 | 0.43 | 0.59 | 0.59 | 0.67 | 1.30 | 0.18 | 0.34 | 0.90 | 1.08 | 1.82 | 1.21 |  |
| 9 | 0.20 | 0.23 | 0.99 | 0.33 | 0.19 | 0.16 | 0.33 | 0.35 | 0.70 | 0.66 | 0.13 | 0.92 | 0.27 |  |
| $10+$ | 0.47 | 0.55 | 0.98 | 0.94 | 1.59 | 0.39 | 1.19 | 0.99 | 1.05 | 0.95 | 0.92 | 1.20 | 0.44 |  |
| Total (ages 1-10+) | 72.12 | 105.18 | 81.96 | 90.17 | 109.64 | 102.73 | 153.13 | 163.47 | 290.15 | 191.52 | 243.43 | 456.19 | 364.37 |  |

Table 2 - Previous index for plaice in the UK-7D BTS (1989-2014)

| Age/Year | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2.31 | 5.16 | 11.75 | 16.53 | 3.22 | 8.33 | 11.32 | 13.20 | 33.15 | 11.38 | 11.30 | 13.19 | 17.91 |
| 2 | 12.13 | 4.86 | 9.06 | 12.54 | 13.40 | 7.46 | 4.06 | 11.90 | 13.48 | 27.30 | 14.10 | 20.96 | 13.02 |
| 3 | 16.63 | 5.76 | 6.98 | 4.19 | 4.96 | 9.17 | 3.00 | 1.30 | 4.22 | 6.99 | 15.90 | 14.39 | 10.00 |
| 4 | 19.94 | 6.70 | 5.30 | 4.17 | 1.75 | 5.56 | 3.67 | 0.70 | 0.65 | 3.12 | 2.90 | 13.81 | 7.12 |
| 5 | 3.30 | 7.53 | 5.43 | 5.57 | 1.89 | 1.95 | 1.49 | 1.30 | 0.34 | 0.32 | 1.00 | 3.48 | 10.94 |
| 6 | 1.48 | 1.76 | 3.20 | 4.88 | 1.57 | 0.77 | 0.58 | 0.90 | 0.32 | 0.22 | 0.20 | 0.87 | 1.95 |
| 7 | 1.32 | 0.65 | 1.22 | 3.44 | 2.05 | 0.90 | 0.59 | 0.40 | 0.24 | 0.15 | 0.10 | 0.57 | 0.53 |
| 8 | 0.54 | 0.97 | 0.99 | 0.66 | 2.78 | 1.83 | 1.32 | 0.30 | 0.21 | 0.11 | 0.30 | 0.18 | 0.30 |
| 9 | 0.30 | 0.75 | 0.06 | 0.49 | 0.39 | 1.24 | 0.82 | 0.40 | 0.17 | 0.05 | 0.10 | 0.43 | 0.19 |
| $10+$ | 1.65 | 0.37 | 1.24 | 0.72 | 0.57 | 0.81 | 0.78 | 2.80 | 1.86 | 0.98 | 0.90 | 1.52 | 0.99 |
| Total (ages 1-10+) | 59.60 | 34.51 | 45.23 | 53.19 | 32.57 | 38.03 | 27.63 | 33.20 | 54.64 | 50.62 | 46.80 | 69.40 | 62.94 |
| Age/Year | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| 1 | 20.66 | 6.18 | 36.18 | 10.84 | 17.21 | 42.61 | 30.28 | 71.62 | 65.25 | 105.55 | 23.23 | 34.33 | 153.63 |
| 2 | 15.95 | 22.79 | 14.97 | 31.21 | 16.11 | 18.81 | 26.52 | 42.88 | 63.83 | 95.31 | 76.07 | 59.27 | 140.96 |
| 3 | 7.73 | 6.00 | 13.15 | 13.77 | 9.22 | 8.70 | 7.20 | 19.15 | 17.27 | 35.70 | 45.26 | 87.99 | 50.67 |
| 4 | 3.55 | 2.94 | 3.44 | 10.28 | 3.35 | 3.87 | 2.97 | 5.74 | 8.90 | 9.25 | 12.73 | 45.47 | 55.50 |
| 5 | 1.80 | 1.61 | 0.91 | 2.95 | 2.64 | 1.75 | 2.32 | 3.20 | 3.04 | 6.68 | 3.53 | 10.58 | 25.08 |
| 6 | 3.46 | 0.79 | 0.16 | 1.17 | 0.77 | 1.95 | 1.11 | 2.17 | 1.90 | 2.82 | 1.61 | 3.54 | 9.13 |
| 7 | 0.72 | 1.77 | 0.66 | 0.77 | 0.57 | 0.80 | 0.50 | 0.78 | 1.38 | 1.40 | 0.42 | 1.03 | 2.32 |
| 8 | 0.14 | 0.60 | 1.16 | 0.42 | 0.31 | 0.30 | 0.41 | 1.24 | 0.30 | 0.19 | 0.41 | 1.37 | 1.88 |
| 9 | 0.11 | 0.11 | 0.17 | 0.86 | 0.14 | 0.10 | 0.09 | 0.37 | 0.36 | 0.57 | 0.43 | 0.14 | 1.01 |
| $10+$ | 0.61 | 0.28 | 0.17 | 0.65 | 0.46 | 1.11 | 0.25 | 1.31 | 0.89 | 0.95 | 0.12 | 0.20 | 1.36 |
| Total (ages 1-10+) | 54.71 | 43.06 | 70.97 | 72.91 | 50.79 | 80.01 | 71.66 | 148.46 | 163.10 | 258.41 | 163.82 | 243.92 | 441.55 |

Table 3 - Revised index for sole in the UK-7D BTS (1989-2015)

| Age/Year | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.16 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.06 | 5.55 | 0.06 | 0.13 | 2.56 | 0.00 | 1.27 | 0.00 |
| 1 | 3.01 | 17.96 | 12.14 | 1.33 | 0.82 | 8.33 | 5.89 | 5.30 | 24.75 | 3.27 | 35.99 | 14.98 | 10.19 | 53.56 |
| 2 | 22.09 | 5.55 | 31.17 | 15.29 | 22.96 | 4.26 | 16.09 | 10.79 | 10.85 | 24.11 | 8.22 | 27.45 | 27.88 | 16.11 |
| 3 | 4.62 | 5.55 | 3.19 | 13.47 | 11.42 | 11.07 | 2.22 | 5.97 | 4.42 | 3.67 | 11.33 | 5.52 | 11.55 | 8.60 |
| 4 | 2.45 | 1.24 | 2.82 | 1.07 | 9.97 | 4.65 | 3.51 | 1.07 | 1.94 | 1.47 | 1.59 | 4.85 | 1.67 | 5.11 |
| 5 | 0.56 | 1.01 | 0.48 | 1.61 | 1.14 | 4.30 | 1.67 | 1.86 | 0.26 | 0.83 | 0.73 | 1.48 | 2.33 | 0.45 |
| 6 | 0.35 | 0.33 | 0.67 | 0.34 | 1.52 | 0.28 | 2.12 | 1.15 | 0.82 | 0.19 | 1.02 | 0.68 | 0.75 | 1.04 |
| 7 | 0.26 | 0.06 | 0.16 | 0.50 | 0.34 | 0.90 | 0.28 | 1.55 | 0.52 | 0.37 | 0.19 | 0.34 | 0.63 | 0.59 |
| 8 | 0.05 | 0.15 | 0.20 | 0.11 | 0.34 | 0.09 | 0.53 | 0.20 | 0.96 | 0.08 | 0.54 | 0.00 | 0.48 | 0.17 |
| 9 | 0.00 | 0.00 | 0.07 | 0.30 | 0.07 | 0.46 | 0.20 | 0.65 | 0.07 | 0.13 | 0.43 | 0.34 | 0.12 | 0.00 |
| $10+$ | 0.72 | 0.16 | 0.26 | 1.11 | 0.40 | 0.46 | 0.32 | 0.59 | 0.62 | 0.35 | 0.54 | 1.06 | 0.86 | 0.72 |
| Total (ages 1-10+) | 34.11 | 32.00 | 51.14 | 35.15 | 48.98 | 34.80 | 32.84 | 29.14 | 45.21 | 34.48 | 60.59 | 56.70 | 56.46 | 86.36 |
| Age/Year | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |  |
| 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.45 | 0.00 |  |
| 1 | 11.03 | 12.67 | 43.27 | 10.84 | 2.57 | 3.77 | 51.25 | 16.59 | 13.66 | 1.75 | 0.72 | 25.39 | 25.24 |  |
| 2 | 45.65 | 11.81 | 6.91 | 42.62 | 28.97 | 7.35 | 19.16 | 30.76 | 28.60 | 9.72 | 8.91 | 16.35 | 21.36 |  |
| 3 | 5.87 | 10.97 | 3.50 | 4.51 | 15.45 | 9.14 | 7.10 | 5.14 | 14.70 | 7.51 | 15.09 | 12.38 | 6.04 |  |
| 4 | 3.20 | 2.08 | 5.18 | 2.68 | 1.47 | 5.82 | 5.81 | 1.66 | 1.66 | 3.53 | 9.72 | 11.92 | 2.29 |  |
| 5 | 2.05 | 2.02 | 1.90 | 2.59 | 1.04 | 0.40 | 5.02 | 2.70 | 0.54 | 0.92 | 3.23 | 5.09 | 4.51 |  |
| 6 | 0.42 | 1.34 | 1.15 | 0.55 | 1.56 | 0.68 | 0.44 | 2.73 | 2.62 | 0.39 | 1.12 | 2.73 | 2.08 |  |
| 7 | 0.55 | 0.41 | 0.71 | 0.47 | 0.44 | 0.37 | 0.31 | 0.33 | 0.77 | 0.78 | 0.51 | 1.08 | 2.20 |  |
| 8 | 0.27 | 0.64 | 0.08 | 0.66 | 0.21 | 0.37 | 0.63 | 0.06 | 0.24 | 0.67 | 0.89 | 0.32 | 0.20 |  |
| 9 | 0.03 | 0.26 | 0.36 | 0.00 | 0.55 | 0.25 | 0.26 | 0.49 | 0.19 | 0.00 | 0.78 | 0.20 | 0.00 |  |
| $10+$ | 0.92 | 0.88 | 0.35 | 0.40 | 0.53 | 0.26 | 0.59 | 0.31 | 0.12 | 0.70 | 0.17 | 0.70 | 0.67 |  |
| Total (ages 1-10+) | 69.99 | 43.08 | 63.40 | 65.32 | 52.79 | 28.41 | 90.58 | 60.78 | 63.11 | 25.97 | 41.13 | 76.15 | 64.60 |  |

Table 4 - Previous index for sole in the UK-7D BTS (1989-2014)

| Age/Year | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2.60 | 12.10 | 8.90 | 1.40 | 0.50 | 4.80 | 3.50 | 3.50 | 19.00 | 2.00 | 28.14 | 10.49 | 9.09 |
| 2 | 15.40 | 3.70 | 22.80 | 12.00 | 17.50 | 3.20 | 10.60 | 7.30 | 7.30 | 21.20 | 9.44 | 22.03 | 21.01 |
| 3 | 3.40 | 3.40 | 2.20 | 10.00 | 8.40 | 8.30 | 1.50 | 3.80 | 3.20 | 2.50 | 13.17 | 4.15 | 8.36 |
| 4 | 1.70 | 0.70 | 2.30 | 0.70 | 7.00 | 3.30 | 2.30 | 0.70 | 1.30 | 1.00 | 2.51 | 4.24 | 1.20 |
| 5 | 0.60 | 0.80 | 0.30 | 1.10 | 0.80 | 3.30 | 1.20 | 1.30 | 0.20 | 0.90 | 1.73 | 1.03 | 1.91 |
| 6 | 0.20 | 0.20 | 0.50 | 0.30 | 1.00 | 0.20 | 1.50 | 0.90 | 0.50 | 0.10 | 1.28 | 0.58 | 0.54 |
| 7 | 0.20 | 0.10 | 0.10 | 0.50 | 0.30 | 0.60 | 0.20 | 1.10 | 0.40 | 0.30 | 0.16 | 0.28 | 0.57 |
| 8 | 0.00 | 0.20 | 0.20 | 0.10 | 0.20 | 0.10 | 0.30 | 0.10 | 0.90 | 0.00 | 0.93 | 0.03 | 0.35 |
| 9 | 0.00 | 0.00 | 0.10 | 0.20 | 0.00 | 0.30 | 0.20 | 0.50 | 0.00 | 0.10 | 1.07 | 0.24 | 0.04 |
| $10+$ | 0.70 | 0.00 | 0.10 | 0.60 | 0.40 | 0.30 | 0.30 | 0.40 | 0.70 | 0.30 | 0.47 | 1.20 | 1.01 |
| Total (ages 1-10+) | 24.80 | 21.20 | 37.50 | 26.90 | 36.10 | 24.40 | 21.60 | 19.60 | 33.50 | 28.40 | 58.89 | 44.28 | 44.09 |
| Age/Year | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| 1 | 31.76 | 6.47 | 7.35 | 25.00 | 6.30 | 2.14 | 2.86 | 30.54 | 15.90 | 11.92 | 1.77 | 0.78 | 25.53 |
| 2 | 11.42 | 28.48 | 8.49 | 5.04 | 29.18 | 21.86 | 6.46 | 13.33 | 30.12 | 23.54 | 9.28 | 9.20 | 13.93 |
| 3 | 5.42 | 4.13 | 7.71 | 2.86 | 2.83 | 12.90 | 7.24 | 5.44 | 5.32 | 11.56 | 6.57 | 15.54 | 9.87 |
| 4 | 3.45 | 2.46 | 1.57 | 3.47 | 1.99 | 1.22 | 4.82 | 4.34 | 1.66 | 1.25 | 3.41 | 8.91 | 11.31 |
| 5 | 0.27 | 1.58 | 1.45 | 1.63 | 1.95 | 0.80 | 0.25 | 3.76 | 2.82 | 0.57 | 0.88 | 2.95 | 5.22 |
| 6 | 0.71 | 0.30 | 0.99 | 1.02 | 0.34 | 1.20 | 0.49 | 0.37 | 2.38 | 2.56 | 0.39 | 1.35 | 3.52 |
| 7 | 0.44 | 0.39 | 0.20 | 0.66 | 0.44 | 0.32 | 0.38 | 0.20 | 0.35 | 0.60 | 0.66 | 0.37 | 1.40 |
| 8 | 0.09 | 0.20 | 0.44 | 0.06 | 0.57 | 0.17 | 0.27 | 0.31 | 0.16 | 0.16 | 0.52 | 0.97 | 0.85 |
| 9 | 0.00 | 0.07 | 0.21 | 0.31 | 0.00 | 0.59 | 0.24 | 0.23 | 0.55 | 0.21 | 0.00 | 0.75 | 0.23 |
| $10+$ | 0.56 | 0.52 | 0.57 | 0.35 | 0.34 | 1.02 | 0.20 | 0.48 | 0.31 | 0.06 | 0.66 | 0.10 | 0.26 |
| Total (ages 1-10+) | 54.12 | 44.60 | 28.98 | 40.40 | 43.93 | 42.22 | 23.21 | 59.01 | 59.56 | 52.44 | 24.16 | 40.92 | 72.11 |



Figure 1 - Prime stations for Q3 UK beam trawl survey for survey index calculation (1989-2015)


Figure 2 - Long-term trends of plaice survey index in the UK - 7D BTS (revised and previous index) for 1-year to 6-year class.


Figure 3 - Long-term trends of sole survey index in the UK - 7D BTS (revised and previous index) for 1-year to 6-year class.

WGNSSK had some concerns about the saithe assessment model.

- Running the forecast with the benchmark-approved model resulted in unrealistically high increases in TAC for the advice ( $119 \%$ increase, MSY approach). The working group asked to review the model with only the standardized combined cpue index tuned to the exploitable biomass.
- A model using both the standardized combined cpue index (FBI) and the IBTS Q3 survey was put forward as an alternate model. Because this model diverges from the cpue-only model, properties of the survey were investigated (e.g., internal consistency, cross-consistency with other data, coverage).
- This prompted a more thorough exploration of the survey data to:
- Determine if spatial changes had occurred in the survey that could be the result of fish moving in and out of the survey area (unrelated to stock size).
- Investigate the Q3 index models.
- Include a ship effect to determine whether a newly added ship at the end of the time series might be causing the problem (e.g., Dana in Skagerrak).
- Modify the spatial grid over which the indices are estimated so that it is roughly representative of the population (do not include large areas where there are almost no saithe). Two potential indices were explored: one that removed the Skagerrak/Kattegat and southern North Sea (south of $57^{\circ} \mathrm{N}$ ); one that kept the Skagerrak but removed the Kattegat and southern North Sea.
- Investigate consistencies for each model option.
- Determine the effect of various age groups.
- There were questions regarding the use of SAM vs. XSA. Discussions via email may have put this option to rest, but are summarized as:
- When XSA and SAM are run with the same datasets, the XSA results fall more-or-less within the confidence limits of SAM (Figure 1).
- Reverting to XSA would actually hamper our ability to investigate the uncertainties arising from the different datasets.
- The 3 cpue indices get a very high weight and the IBTS q3 has hardly any influence; this hides the issue that the assessment relies nearly entirely on the commercial indices.
- Would need to revert back to the age-based cpue indices because XSA cannot handle the combined standardized index, that is fit to the exploitable biomass (within the model). This reverts back to the issue of using the age information twice - once for the catch data, once for the cpue tuning indices.
- The XSA cannot handle the correlation between ages with years in the survey indices; SAM can, as outlined in Berg \& Nielsen (2016).

A bug in InterCatch resulted in the re-raising of discards for 2003, 2006, 2011, and 2014, which were done following the procedure in WD 5; 2002 was also re-raised as it seemed oddly high. After re-raising the data, several years still appeared to be atypical, so the
raising for all years was re-done following a modification to the rules used for the benchmark:

- No discard ratio $>=25 \%$ was used in the raising of any fleet. Previously, ratios $>30 \%$ were omitted.
- Norwegian trawler fleet discards were raised using German or French (or both) discard information. Previously, they were raised with other OTB_DEF fleets, using discard information from all OTB_DEF fleets for a given area and quarter.


## Results

## Spatial changes in the surveys

Spatial plots of the catches (all ages combined) showed that, for the Q1 survey, saithe were mainly on the shelf edges and the survey was unlikely to be sampling much of the population (see Appendix: Q1 plots are catch weight per station per year, not age specific). At the time of the benchmark, this was discussed, but it was thought that, for the older ages, the amount of the population surveyed should be fairly consistent over time. A month parameter had been added to the delta-GAM model to account for changes in survey timing and any effect of fish movement in and out of the survey area. However, closer inspection of the figures showed that, in some years, fish are found further up on the shelf, while in other years, they are only along the shelf boundary ( 200 m contour). This does call into question using the Q1 index in the assessment.

For the Q3 surveys, saithe are found on the northern part of the shelf, along the shelf boundary, and in the Skagerrak (see Appendix: Q3 plots are catch weight per station per year, not age specific). The amount of saithe found within the area differs, but the distribution appeared fairly consistent. Stronger year classes are, for the most part, appearing in the survey when expected and persisting for at least 1 year (e.g., 1995, 2001, 2005).

## Q3 index models

A ship effect was included in the index estimation. Sweden had begun using a new vessel in 2011 in the Skagerrak. Including ship in the model resulted in a higher AIC and BIC, and slightly worse internal consistencies (Tables 1, 2).

The spatial grid was truncated to a) exclude the area east of $8^{\circ} \mathrm{E}$ and south of $57^{\circ} \mathrm{N}$, i.e., Skagerrak/Kattegat and southern North Sea information were removed, and b) exclude south of $57^{\circ} \mathrm{N}$ and the Kattegat (but include the Skagerrak). Saithe are not found in the southern North Sea; excluding this area mainly truncates the zeros and keeps the spatial spline of the GAM model from attempting to put fish where they are typically not found. Mainly young fish (the ages not included in the assessment model) are found in the Skagerrak, but the German fleet fishes in this area; therefore, datasets including and excluding this region were trialed. Ship was included in the final model.

Truncating the spatial area improved the model fit (Table 1). Removing the Skagerrak improved the fit of the model the most, but the indices were larger for a given age class and more variable for many of the age classes, especially at the beginning of the time series (Figure 2). Average internal consistencies were higher for the model including the Skagerrak, but the fit was not as good as the model excluding the Skagerrak (Tables 1,2). Figure 3 shows the internal consistency plot, as given by FLR (note: correlations are reported differently using FLR); there is no evidence in the internal consistencies
that something has gone wrong in the survey. The time series of indices by age (including confidence intervals and comparison to the DATRAS indices) for the full survey area, excluding the southern North Sea and Skagerrak/Kattegat, and excluding the southern North Sea and Kattegat are in Figures 4-6.
The effect seen at the start of the time series cannot be due to ship; it would have been captured within the model or also seen from 2001, when Sweden changed its research vessel. The indices (all ages) with and without the Skagerrak show similar trends and values.

Until 2003, Sweden did not take age samples, only lengths. This resulted in the agelength key for the North Sea (subarea 4) being applied to the Skagerrak. Whether fish in the Skagerrak were different from the North Sea was not thoroughly investigated, so it is questionable whether the age-length key from the North Sea should be applied to the Skagerrak. In addition, Sweden did not survey in 2000; this year had incomplete coverage of the entire survey area. Finally, the Skagerrak was never included in the old index estimation (in DATRAS). There is no documentation of why the Skagerrak was included and the IBTSWG was unable to answer this question.

## Survey properties

Internal consistencies
Internal consistencies for the Q3 survey are decent, although slightly poorer for age 3 vs. age 4 (Table 2, Figure 3). There is no evidence in the internal consistencies that something has gone wrong in the survey.

## Cross consistency with other data sources

Despite the Q1 survey having limited coverage of the stock, the external consistencies between the Q3 and Q1 (in the following year and age), as well as catch numbers at age, were used to see if tracking of cohorts was possible (Table 3). Cohorts can be tracked between surveys (and ages). The external consistencies are not as strong when comparing the catch numbers at age with the Q3 index, however, they still track cohorts reasonably well. The external consistency for age 4 , the age when fish are expected to be fully recruited to the fishery, is the lowest of all the age class comparisons.

## Coverage

The amount of saithe found within the survey area differs between years, but the distribution has not changed over the time period. Stronger year classes are, for the most part, appearing in the survey when expected and persisting for at least 1 year. The increase in the last 2 years appears to be related to stronger recruitment.

## Effect of age groups and research surveys on the assessment

Only the Q3 index was used to assessing the influence of the different age classes. The decision was made that it is not appropriate to continue to include the Q1 indices in the assessment model (see above).

## Q3 indices without truncating spatial grid or including ship in the model

The assessment results when including only the Q3 + FBI indices show SSB in the final years not as optimistic as the model including the Q1 index (Figure 7). It is, however, much more optimistic than the FBI-only model or the model using the DATRAS-estimated indices for ages 3-5. When looking at the effect of removing the oldest age classes one at a time, ages 5-8 have the largest effect on the assessment outcome (Figure 8). Using only the age ranges 3-4 or 3-9 has a large effect on the estimated SSB; ages 3-9 result in a lower SSB over the entire time series, while using only ages 3-4 has a mixed effect (lower SSB after 2010). The effect of changing the age range on $F_{b a r}$ and recruitment are shown in Figure 9.

Q3 indices with truncation of spatial grid + including ship in model
Figure 10 shows the effect of the Q3 (without Skagerrak) index on assessment model outputs. SSB and F are much closer to the DATRAS outputs and below that of the previous Q3 indices. Figures 11 and 12 detail the effects of changing the age range included in the Q3 index on SSB, $\mathrm{F}_{4-7}$, and recruitment.

The effect of the Q3 with Skagerrak indices on the assessment model are in Figure 10. Including the Skagerrak in the Q3 index resulted in output that was similar to the model using the Q3 indices estimated from the entire North Sea dataset (Q3 + FBI model). Figures 13 and 14 show the effect of changing the age range included in the model on SSB, $\mathrm{F}_{4-7}$, and recruitment.

## Discard estimation

The change in discard amounts are in Table 4. The years that had the greatest percentage difference due to the modifications noted above were the years that had very few reported discards; Norwegian discards had to be estimated using poor data. Norway takes $50 \%$ of the catch and this therefore resulted in high raised discards amounts. Because there is no information on the discarding practices of Norwegian fleets, the truth is expected to lie somewhere between estimate (3) and estimate (2); these estimates should be treated as upper and lower bounds on discards. It is doubtful that Norwegian discards are at the low levels estimated in option 3. However, when low recruitment is seen (2008-2010), discards should be low. This is seen in Table 3 using raising option (3), but not in option (2). While raising option (3) may be under-estimating discards, it appears to be more likely than option (2).

The comparison of assessments (old raising procedure vs. option (3)) for the benchmark model (FBI + Q1 + Q3), FBI index-only model, and new Q3 model, where the Skagerrak/Kattegat/southern North Sea were truncated from the spatial grid are in Figures 15-17. Results of all 3 models using revised discards data are in Figure 18.
Retrospectives using the newly estimated catch are in Figure 19 for the benchmark model (FBI + Q1 + Q3) without discard revisions. Figure 20 is the benchmark model including discard revisions, Figure 21 is the FBI-only model (including discards revisions), and Figure 22 for the FBI+ new Q3 model (including discard revisions). The retrospective pattern is much worse for the benchmark model with the revised catch
information. The retrospective pattern in F is particularly bad. The model with only the exploitable biomass index shows the best performance in the retrospective analysis.
All models converge to approximately similar F and SSB values for the 2005-2010 period (Figure 18). Therefore, by going back with the retro analysis before 2010 gives an idea which assessment would have been more in line with the final converged values. The assessment with FBI as only index would have assessed F around the converged values for 2005-2010. The retrospective indicates all other models would have assessed F well above the converged values for this period (with the FBI + new Q3 model being the worst)In recent years the retro patterns became less, however each of the assessments show F at a different level. It remains unclear whether the current FBI only assessment will be again closer to the converged estimates in a few years. Reference points and catch option tables are in the Appendix for the 3 models with revised catch information.

## Conclusions

The Q1 index should not be included as a tuning series because the survey does not adequately cover the distribution of saithe. Saithe are spawning on the slope and their movement into (or out of) the survey area does not appear to be linked to recruitment or expected abundance.

For the Q3 index, the spatial distribution of saithe has not changed within the survey area. Truncating the spatial grid to be remove the southern North Sea (where saithe are not found) and the Skagerrak should be done. The arguments for excluding the Skagerrak include: no age-length key in the Skagerrak until after 2003, incomplete coverage of the survey area due to Skagerrak not surveyed in 2000, and exclusion of the Skagerrak in the previous (DATRAS) index estimation (even though the reason is not known).

Removing the Skagerrak and southern North Sea resulted in a less optimistic assessment when compared to the benchmark model. The assessment using Q3 indices that included the Skagerrak, but removing the Kattegat and southern North Sea, was (not surprisingly) similar to the benchmark assessment. The data from the Skagerrak appears to be creating an issue with the index estimation. The reason for this is not clear (biological or a survey effect, due to the lack of age information from this area). The reason for the large discrepancy in the indices including/excluding the Skagerrak for the beginning of the series should be investigated in the near future.
Because Norway lacks information on discards and takes $50 \%$ of the catch, the raising of discards in InterCatch must be handled carefully. Raising discards for the Norwegian trawlers based on reported discards from the French and German trawlers may result in underestimating the discards, but it is the best information available at this time. Germany, France and Norway have a targeted saithe fishery. Fisheries in countries like Scotland and Denmark are mixed demersal fisheries with higher discard rates compared to the sampled fisheries targeting saithe.

The pre-benchmark assessment included the Q3 indices for ages 3-5. The internal consistencies, coverage, and comparison with other data all show no reason to exclude the survey from the assessment. It is only in the last two years that the assessment has shown SSB is higher than the cpue-only model; prior to 2013, the cpue-only model had consistently higher SSB (Figure 18). There is a lot of uncertainty in the assessment regardless of the model chosen. The choice of survey data to include should be based on the properties of that survey (e.g., internal consistency, cross-consistency with other data, coverage).

The retrospective patterns, particularly for F , were very poor, especially for the assessments with IBTS data included. This is worrying as it casts doubt on our ability to assess the stock should conditions change again. Furthermore, the cause for the poor ability to estimate F is unknown (and could occur again). There is some doubt that the FBI + new Q3 model is the better model compared to the FBI-only model in light of the retrospective patterns.

Keeping the stipulation from the EU-Norway management plan, where the TAC is not allowed to deviate by more than $15 \%$ from the TAC in the previous year should protect the stock from the uncertainty in the assessment. Furthermore, including catch options based on probabilistic forecasts, e.g., $5 \%$ and $25 \%$ probability of being above $\mathrm{F}_{\text {MSY }}$ and Flim, is another option for dealing with the uncertainty in the assessment.

Table 1. Model diagnostics for the Q3 indices. The models are the benchmark model (no truncation of spatial grid); benchmark model including Ship (no truncation of spatial grid); removing the Skagerrak/Kattegat and southern North Sea and including Ship; removing the Kattegat and southern North Sea and including Ship.

| Model | AIC | BIC | IC (All ages) |
| :---: | :---: | :---: | :---: |
| Year+s(lon,lat)+s(Depth)+ HaulDur | 34460 | 42834 | 0.3948 |
| Year+Ship+s(lon,lat)+s(Depth)+HaulDur | 34274 | 43476 | 0.4358 |
| Truncated spatial range ( $57^{\circ} \mathrm{N}, 8^{\circ} \mathrm{E}$ ): <br> Year+Ship+s(lon,lat)+s(Depth)+HaulDur, ages 110 | 28122 | 36032 | 0.40527 |
| Truncated spatial range $\left(57^{\circ} \mathrm{N}\right.$, no Kattegat): Year+Ship+s(lon,lat)+s(Depth)+HaulDur, ages 110 | 32565 | 40590 | 0.4264 |

Table 2. Internal consistencies between ages classes for the four different Q3 indices.

| Model/Data | IC | Average IC ALL AGES | Average IC AGES 3-8 |
| :---: | :---: | :---: | :---: |
| Benchmark model: | Age 0 vs. $1: 0.3231104$ <br> Age 1 vs. 2 : -0.1937066 | 0.3948 | 0.6851 |
| $\begin{aligned} & \text { Year+s(lon,lat)+s(Depth)+ HaulDur, } \\ & \text { ages 0-10 } \end{aligned}$ | Age 2 vs. 3 : 0.03960032 <br> Age 3 vs. $4: 0.4954253$ <br> Age 4 vs. $5: 0.7447504$ <br> Age 5 vs. $6: 0.7943942$ <br> Age 6 vs. $7: 0.750217$ <br> Age 7 vs. $8: 0.6407721$ <br> Age 8 vs. $9: 0.4044236$ <br> Age 9 vs. $10:-0.05130193$ |  |  |
| Benchmark model + Ship: <br> Year+Ship $+s($ lon,lat $)+s($ Depth $)+$ HaulDur, ages 0-10 | Age 0 vs. $1: 0.4646722$ <br> Age 1 vs. 2 : -0.1081123 <br> Age 2 vs. $3: 0.05023302$ <br> Age 3 vs. $4: 0.4406976$ <br> Age 4 vs. $5: 0.7406408$ <br> Age 5 vs. $6: 0.8363853$ <br> Age 6 vs. $7: 0.7676941$ <br> Age 7 vs. $8: 0.5378916$ <br> Age 8 vs. $9: 0.3850141$ <br> Age 9 vs. $10: 0.2426996$ | 0.4358 | 0.6647 |
| Truncated spatial range (no Skagerrak/Kattegat or southern North Sea): Year+Ship+s(lon,lat)+s(Depth)+HaulDur, ages 1-10 | Age 1 vs. $2: 0.4287579$ <br> Age 2 vs. 3 : 0.1669562 <br> Age 3 vs. $4: 0.3777139$ <br> Age 4 vs. $5: 0.759958$ <br> Age 5 vs. $6: 0.7629555$ <br> Age 6 vs. $7: 0.7211942$ <br> Age 7 vs. $8: 0.6095779$ <br> Age 8 vs. 9 : 0.08241081 <br> Age 9 vs. $10:-0.262115$ | 0.4053 | 0.6463 |
| Truncated spatial range (no Kattegat or southern North Sea): <br> Year+Ship+s(lon,lat)+s(Depth)+HaulDur, ages 1-10 | Age 1 vs. 2 : -0.3264555 <br> Age 2 vs. $3:-0.02738525$ <br> Age 3 vs. $4: 0.4273828$ <br> Age 4 vs. $5: 0.7532319$ <br> Age 5 vs. $6: 0.8270072$ <br> Age 6 vs. $7: 0.7994671$ <br> Age 7 vs. $8: 0.6195003$ <br> Age 8 vs. $9: 0.3514482$ <br> Age 9 vs. $10: 0.4138057$ | $0.4264$ | 0.6853 |

Table 3. External consistencies between Q3 (ages 1-9, 1992-2014) and Q1 (year+1, age+1), and between catch numbers at age and Q3 (in the same year). This is identifying if cohorts can be tracked from Q3 to the next survey in Q1. Numbers in bold refer to the ages included in the IBTS Q3 tuning index in the assessment model.

| External consistencies: Q3 vs. Q1 | External consistencies: CATCH vs. Q3 |
| ---: | :--- |
| Q3 Age 1 vs. Q1 Age $2: 0.3218696$ | Catch Age 1 vs. Q3 $1:-0.0165032$ |
| Q3 Age 2 vs. Q1 Age $3: 0.4586471$ | Catch Age 2 vs. Q3 $2:-0.1492027$ |
| Q3 Age 3 vs. Q1 Age $4: 0.8203473$ | Catch Age 3 vs. Q3 $3: 0.5044318$ |
| Q3 Age 4 vs. Q1 Age $5: 0.8739198$ | Catch Age 4 vs. Q3 $4: 0.3768049$ |
| Q3 Age 5 vs. Q1 Age $6: 0.8839688$ | Catch Age 5 vs. Q3 $5: 0.5894862$ |
| Q3 Age 6 vs. Q1 Age $7: 0.7743481$ | Catch Age 6 vs. Q3 $6: 0.5557922$ |
| Q3 Age 7 vs. Q1 Age $8: 0.696888$ | Catch Age 8 vs. Q3 $8: 0.4097457$ |
| Q3 Age 8 vs. Q1 Age $9: 0.625716$ | Catch Age 9 vs. Q3 $9: 0.05872988$ |
| Q3 Age 9 vs. Q1 10 : 0.4047001 | Catch Age 10 vs. Q3 $10: 0.4557988$ |

Table 4. Amount of discards (estimated and reported) following 3 procedures: (1) as outlined in WD-5 during the benchmark, (2) after fixing the bug in InterCatch (bolded years), and (3) after the modification noted above. Differences are percentage.

| (3) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Modification |  |  |  |  |
|  |  |  | (2) | to Norway \& |  |  |  |  |
|  |  | (1) | InterCatch | Reduced |  | Difference |  |  |
|  | 2015 | Benchmark | BUG | RATIO |  | 2015 то | Difference | Difference |
| Year | ASSESSMENT | ESTIMATE | CORRECTION | EStimate | Reported | (1) | (1) то (2) | (2) то (3) |
| 2002 |  | 24812 | 21620 | 21544 | 21440 | 100 | -13 | 0 |
| 2003 |  | 26377 | 12898 | 11438 | 11044 | 100 | -51 | -11 |
| 2004 |  | 9600 | 9656 | 8088 | 7850 | 100 | 1 | -16 |
| 2005 |  | 8571 | 8571 | 8196 | 8072 | 100 | 0 | -4 |
| 2006 |  | 15950 | 9498 | 8585 | 8340 | 100 | -40 | -10 |
| 2007 |  | 12050 | 12078 | 12413 | 11353 | 100 | 0 | 3 |
| 2008 |  | 9436 | 9436 | 8359 | 7891 | 100 | 0 | -11 |
| 2009 |  | 14216 | 14216 | 4296 | 4170 | 100 | 0 | -70 |
| 2010 |  | 10937 | 10937 | 4484 | 3009 | 100 | 0 | -59 |
| 2011 |  | 12729 | 4951 | 4362 | 4285 | 100 | -61 | -12 |
| 2012 | 7585 | 9415 | 9415 | 9278 | 7471 | 24 | 0 | -1 |
| 2013 | 8083 | 8173 | 8173 | 7777 | 7311 | 1 | 0 | -5 |
| 2014 | 6289 | 6362 | 6356 | 6337 | 6068 | 1 | 0 | 0 |
| 2015 |  | 5060 | 5060 | 5003 | 4914 |  | 0 | -1 |



Figure 1. Comparison of the 2015 assessments. Blue lines: XSA assessment results. Black lines: 95\% confidence interval of SAM assessment.


Figure 2. Comparison of IBTS Q3 indices, 1992-2015. Black lines: benchmark Q3 indices (no spatial truncation, without 'Ship' in model); blue lines: truncated spatial grid (No Skagerrak) + 'Ship' in model; red lines: truncated spatial grid (including Skagerrak) + 'Ship' in model.


Figure 3. Internal consistencies as given by FLR. Note: FLR internal consistencies are estimated differently from Berg et al. 2014, as given in the amendment to WD 8.


Figure 4. IBTS Q3 indices, ages $\mathbf{0 - 1 0}, 1992-2015$. Comparing survey indices by age (and confidence interval) to DATRAS indices for the full spatial range-no ship delta-GAM model (Q3 index as presented in the benchmark and WGNSSK).


Figure 5. IBTS Q3 indices, ages 1-10+, 1992-2015. Comparing survey indices by age (and confidence interval) to DATRAS indices (ages 1-6+) for the truncated spatial grid-with ship delta-GAM model; this data excludes the Skagerrak-Kattegat and southern North Sea.


Figure 6. IBTS Q3 indices, ages 1-10+, 1992-2015. Comparing survey indices by age (and confidence interval) to DATRAS indices (ages 1-6+) for the truncated spatial grid-with ship delta-GAM model; this data includes the Skagerrak and excludes the southern North Sea and Kattegat.


Figure 7. Affect of different indices on SAM assessment: black line $=$ benchmark model (Q3 + Q1 + FBI indices); blue line = FBI index only (no surveys); green line = Q3 + FBI indices (no Q1); orange line $=$ DATRAS Q3 (ages 3-5) + FBI indices. The Q3 indices estimated from the delta-GAM are those used in the benchmark meeting (no truncation of the survey area, without Ship in the model). Note: this was conducted before the bug in InterCatch was found and discards had not been re-raised.


Figure 8. Effect of adding an changing age range of the Q3 index on estimated SSB. The Q3 indices were estimated using data from the entire North Sea. Note: this was conducted before the bug in InterCatch was found and discards had not been re-raised.


Figure 9. Effect of adding an changing age range of the Q3 index on estimated (left) $\mathrm{F}_{\text {bar 4-7 }}$ and (right) recruitment. Note: this was conducted before the bug in InterCatch was found and discards had not been re-raised.


Figure 10. Affect of different indices on SAM assessment: black line = benchmark model (Q3 + Q1 + FBI indices); blue line = FBI index only (no surveys); green line = bm_Q3 + FBI indices (no Q1); orange line = DATRAS Q3 (ages 3-5) + FBI indices; magenta line $=$ new Q3 + FBI indices (no Q1); brown line = Q3 including Skagerrak (without Kattegat or southern North Sea) + FBI. The bm_Q3 indices are those used in the benchmark meeting (no truncation of the survey area, without Ship in the model), while the new Q3 indices include truncating the spatial grid + ship in the delta-GAM model. Note: this was conducted before the bug in InterCatch was found and discards had not been re-raised.


Figure 11. Effect of adding an changing age range of the Q3 index (truncated to remove the southern North Sea and Skagerrak/Kattegat) on estimated SSB. Model bm_ includes the Q3 indices estimated without truncation of the survey area or ship in the model. Note: this was conducted before the bug in InterCatch was found and discards had not been re-raised.


Figure 12. Effect of adding an changing age range of the Q3 indices (truncated to remove the southern North Sea and Skagerrak/Kattegat) on estimated (left) $F_{\text {bar 4-7 }}$ and (right) recruitment. Model bm_includes the Q3 indices estimated without truncation of the survey area or ship in the model. Note: this was conducted before the bug in InterCatch was found and discards had not been reraised.


Figure 13. Effect of adding an changing age range of the Q3 indices (truncated to exclude the southern North Sea and Kattegat) on estimated SSB. Model bm_includes the Q3 indices estimated without truncation of the survey area or ship in the model. Note: this was conducted before the bug in InterCatch was found and discards had not been re-raised.


Figure 14. Effect of adding an changing age range of the Q3 indices (truncated to exclude the southern North Sea and Kattegat) on estimated (left) Fbar 4-7 and (right) recruitment. Model bm_includes the Q3 indices estimated without truncation of the survey area or ship in the model. Note: this was conducted before the bug in InterCatch was found and discards had not been re-raised.


Figure 15. Effect of raising discards under assumption that Norway has low to zero discarding. Comparison of benchmark model (Q1 + Q3 + FBI) before and after changing raising procedure.


Figure 16. Effect of raising discards under assumption that Norway has low to zero discarding. Comparison of FBI index only model (no surveys) before and after changing raising procedure.


Figure 17. Effect of raising discards under assumption that Norway has low to zero discarding. Comparison of new Q3 model (FBI + Q3 - spatial truncation excludes Skagerrak/Kattegat and southern North Sea) before and after changing raising procedure.


Figure 18. Trends in SSB, F47, and recruitment for the 3 models. Blue line: Q1 + Q3 + cpue index model; green line: cpue-only model; black line: Q3 + cpue model; orange/tan shaded region: 95\% confidence interval for the Q3 + cpue model; solid grey line (dashed): old Q3 + cpue model ( $95 \%$ confidence interval). The old Q3 model was estimated without removing the southern North Sea (where saithe are not found) and the Skagerrak (see amendment to WD 8 for details).


Figure 19. Five year retrospective pattern in SSB, $\mathrm{F}_{4-7}$, and recruitment. Model is FBI + Q1 + Q3 (untruncated spatial area) and does not include the discard revisions.


Figure 20. Eight year retrospective pattern in SSB, $\mathrm{F}_{4-7}$, and recruitment. Model is FBI + Q1 + Q3 (untruncated spatial area) and includes the discard revisions.


Figure 21. Eight year retrospective pattern in SSB, F47, and recruitment. Model is FBI index only (no surveys) and includes the discard revisions.


Figure 22. Eight year retrospective pattern in SSB, F $_{4-7}$, and recruitment. Model is FBI + new Q3 (excludes Skagerrak/Kattegat and southern North Sea) and includes the discard revisions.

## APPENDIX

## Reference Points and catch options

Reference Points estimated for the benchmark model, which includes the Q1, Q3 (untruncated spatially), and FBI indices are in Table A1; catch options are in Table A2 and basis for the catch options are in Table A3.

For the model that has only the FBI index (no surveys), reference points are in Table A4, catch options are in Table A5, and basis for the catch options are in Table A6.

Table A7 contains reference points estimated from the assessment model that includes the FBI and spatially truncated Q3 indices, where the Q3 indices do not include the southern North Sea or Skagerrak/Kattegat. Catch options are in Table A8 and basis for the catch options are in Table A9.

Table A1. Reference points estimated using the benchmark model (FBI + Q1+ Q3 no spatial truncation).

| STock |  |
| :--- | :--- |
| Reference point | Value |
| Blim | 115000 |
| Bpa (1.4) | 161000 |
| Bpa (sigma) | 142000 |
| Btrigger | 182000 |
| Flim | 0.55 |
| Fpa (1.4) | 0.393 |
| Fpa (sigma) | 0.419 |
| FMSY without Btrigger | 0.359 |
| FMSY lower without Btrigger | 0.204 |
| FMSY upper without Btrigger | 0.492 |
| New FP.05 (5\% risk to Blim without Btrigger) | 0.422 |
| FMSY upper precautionary without Btrigger | 0.393 |
| FP.05 (5\% risk to Blim with Btrigger) | 0.534 |
| FMSY with Btrigger | 0.396 |
| FMSY lower with Btrigger | 0.209 |
| FMSY upper with Btrigger | 0.694 |
| FMSY upper precautionary with Btrigger | 0.393 |
| MSY (without HCR) | 91480 |
| Median SSB at FMSY (without HCR) | 220827 |
| Median SSB lower precautionary (median at | 195709 |
| FMSY upper precautionary; without HCR) | 420907 |
| Median SSB upper (median at FMSY lower; <br> without HCR) |  |

Sigma $(F)=0.1653818$, sigma $(S S B)=0.1300388$.

Table A2. Saithe in Subareas 4 and 6 and Division 3a. The catch options. All weights in tonnes.

|  |  |  | Wanted |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | CATCH | Wanted |  | F | F |  |  | \% TAC |
|  | Total | Wanted | 3A \& 4 | CATCH |  | (total | (WANTED |  | \% SSB | Change |
|  | Catches | CATCH | 2017 | 62017 |  | CATCH) | CATCH) | SSB | CHANGE | WANTED |
| Rationale | 2017* | 2017* | ** | ** | BASIS | 2017 | 2017 | 2018 | *** | CATCH^ |
| MSY |  |  |  |  |  |  |  |  |  |  |
| approach | 133332 | 127631 | 115634 | 11997 | FMSY | 0.36 | 0.34 | 322434 | -3 | 86 |
| EU-Norway |  |  |  |  | Paragraph 5 of |  |  |  |  |  |
| management |  |  |  |  | management |  |  |  |  |  |
| strategy | 82261 | 78824 | 71415 | 7409 | strategy | 0.21 | 0.20 | 374902 | 13 | 15 |
| Precautionary |  |  |  |  | $\mathrm{SSB}=\min \{1$; |  |  |  |  |  |
| approach | 298993 | 284409 | 257675 | 26734 | SSB2017/Btrigger\} | 1.08 | 1.03 | 161000 | -51 | 315 |
| Zero catch | 0 | 0 | 0 | 0 | $\mathrm{F}=0$ | 0 | 0 | 461461 | 39 | -100 |
| Other options | 79730 | 76405 | 69223 | 7182 | F2016 | 0.20 | 0.19 | 377503 | 14 | 11 |
|  | 71619 | 68601 | 62153 | 6448 | TAC2016 | 0.18 | 0.17 | 386190 | 17 | 0 |
|  | 143773 | 137641 | 124703 | 12938 | Fpa | 0.39 | 0.38 | 311808 | -6 | 101 |

Table A3. Saithe in Subareas 4 and 6 and Division 3a. The basis for the catch options.

| VARIABLE | VALUE | Source | Notes |
| :--- | :--- | :--- | :--- |
| F ages 4-7 (2016) | 68601 t | ICES <br> $(2016 a)$ | TAC constraint (F=0.20) |
| SSB (2016) | 284887 t | ICES <br> $(2016 a)$ | SSB in the intermediate year |
| SSB (2017) | 331048 t | ICES <br> $(2016 a)$ | SSB at the beginning of the TAC year |
| Rage3 (2016) | 101 billion | ICES <br> $(2016 a)$ | Median recruitment resampled from 2003- <br> 2015 |
| Rage3 (2017) | 101 billion | ICES <br> $(2016 a)$ | Median recruitment resampled from 2003- <br> 2015 |
| Total catch (2016) | 71775 t | ICES <br> $(2016 a)$ | Assuming 2015 landings fraction by age |
| Commercial | 68601 t | ICES <br> $(2016 a)$ | TAC 2015 |
| landings (2016) | 3174 t | ICES <br> $(2016 a)$ | Assuming 2015 discard fraction by age |
| Discards (2016) |  |  |  |

Table A4. Reference points estimated using the FBI index only model (no surveys).

| STock |  |
| :--- | :--- |
| Reference point | Value |
| Blim | 115000 |
| Bpa (1.4) | 161000 |
| Bpa (sigma) | 151000 |
| Btrigger | 161000 |
| Flim | 0.506 |
| Fpa (1.4) | 0.361 |
| Fpa (sigma) | 0.364 |
| FMSY without Btrigger | 0.361 (was 0.405) |
| FMSY lower without Btrigger | 0.208 |
| FMSY upper without Btrigger | 0.454 |
| New FP.05 (5\% risk to Blim without Btrigger) | 0.384 |
| FMSY upper precautionary without Btrigger | 0.361 |
| FP.05 (5\% risk to Blim with Btrigger) | 0.447 |
| FMSY with Btrigger | 0.38 |
| FMSY lower with Btrigger | 0.211 |
| FMSY upper with Btrigger | 0.595 |
| FMSY upper precautionary with Btrigger | 0.361 |
| MSY (without HCR) | 82466 |
| Median SSB at FMSY (without HCR) | 197952 |
| Median SSB lower precautionary (median at | 197952 |
| FMSY upper precautionary; without HCR) | 377258 |
| Median SSB upper (median at FMSY lower; |  |
| without HCR) |  |

Sigma $(F)=0.1651571$, sigma $(S S B)=0.1997418$.

Table A5. Saithe in Subareas 4 and 6 and Division 3a. The catch options. All weights in tonnes.

|  | Wanted |  |  |  |  | F F |  |  | \% TAC |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | CATCH | Wanted | BASIS |  |  |  |  |  |
|  | Total | Wanted | 3A \& 4 | CATCH |  | (total | (WANTED |  | \% SSB | Change |
|  | CATCHES | CATCH | 2017 | 62017 |  | CATCH) | CATCH) | SSB | CHANGE | WANTED |
| Rationale | 2017 * | 2017* | ** | ** |  | 2017 | 2017 | 2018 | *** | CATCH^ |
| MSY |  |  |  |  |  |  |  |  |  |  |
| EU-Norway management strategy | 84592 | 79134 | 71695 | 7439 | Paragraph 5 of management strategy | 0.33 | 0.31 | 228215 | 8 | 15 |
| Precautionary approach | 157658 | 147368 | 133515 | 13853 | $\begin{aligned} & \text { SSB }=\min \{1 ; \\ & \text { SSB2017/Btrigger }\} \end{aligned}$ | 0.71 | 0.68 | 161000 | -24 | 115 |
| Zero catch | 0 | 0 | 0 | 0 | $\mathrm{F}=0$ | 0 | 0 | 309384 | 47 | -100 |
| Other options | 80442 | 75253 | 68179 | 7074 | F2016 | 0.31 | 0.30 | 232171 | 10 | 10 |
|  | 73037 | 68601 | 62153 | 6448 | TAC2016 | 0.28 | 0.26 | 240108 | 14 | 0 |
|  | 91749 | 85822 | 77755 | 8067 | Fpa | 0.36 | 0.34 | 221501 | 5 | 25 |

Table A6. Saithe in Subareas 4 and 6 and Division 3a. The basis for the catch options.

| Variable | Value | Source | Notes |
| :---: | :---: | :---: | :---: |
| F ages 4-7 (2016) | 68601 t | ICES <br> (2016a) | TAC constraint ( $\mathrm{F}=0.31$ ) |
| SSB (2016) | 199173 t | ICES <br> (2016a) | SSB in the intermediate year |
| SSB (2017) | 211158 t | ICES <br> (2016a) | SSB at the beginning of the TAC year |
| Rage3 (2016) | 103 billion | ICES <br> (2016a) | Median recruitment resampled from 20032015 |
| Rage3 (2017) | 103 billion | ICES <br> (2016a) | Median recruitment resampled from 20032015 |
| Total catch (2016) | 72518 t | ICES <br> (2016a) | Assuming 2015 landings fraction by age |
| Commercial <br> landings (2016) | 68601 t | ICES <br> (2016a) | TAC 2015 |
| Discards (2016) | 3917 t | ICES <br> (2016a) | Assuming 2015 discard fraction by age |

Table A7. Reference points estimated using the FBI + Q3 (spatially truncated, excludes southern North Sea and Skagerrak/Kattegat) model.

| Reference point |  |
| :--- | :--- |
| Blim | 121000 |
| Bpa (1.4) | 169000 |
| Bpa (sigma) | 155000 |
| Btrigger | 170000 |
| Flim | 0.514 |
| Fpa (1.4) | 0.367 |
| Fpa (sigma) | 0.376 |
| FMSY without Btrigger | 0.363 |
| FMSY lower without Btrigger | 0.205 |
| FMSY upper without Btrigger | 0.461 |
| New FP.05 (5\% risk to Blim without Btrigger) | 0.394 |
| FMSY upper precautionary without Btrigger | 0.367 |
| FP.05 (5\% risk to Blim with Btrigger) | 0.455 |
| FMSY with Btrigger | 0.382 |
| FMSY lower with Btrigger | 0.208 |
| FMSY upper with Btrigger | 0.607 |
| FMSY upper precautionary with Btrigger | 0.367 |
| MSY (without HCR) | 87658 |
| Median SSB at FMSY (without HCR) | 209632 |
| Median SSB lower precautionary (median at | 206489 |
| FMSY upper precautionary; without HCR) | 402573 |
| Median SSB upper (median at FMSY lower; |  |
| without HCR) |  |

Sigma $(F)=0.189643$, sigma $(S S B)=0.1512602$.

Table A8. Saithe in Subareas 4 and 6 and Division 3a. The catch options. All weights in tonnes.

|  | Wanted |  |  |  |  | F F |  |  | \% TAC |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | CATCH | Wanted |  |  |  |  |  |  |
|  | Total | Wanted | 3A \& 4 | CATCH |  | (TOTAL | (WANTED |  | \% SSB | Change |
|  | CATCHES | CATCH | 2017 | 62017 |  | CATCH) | CATCH) | SSB | CHANGE | WANTED |
| Rationale | 2017* | 2017* | ** | ** | BASIS | 2017 | 2017 | 2018 | *** | CATCH^ |
| MSY |  |  |  |  |  |  |  |  |  |  |
| EU-Norway management strategy | 82766 | 78909 | 71492 | 7417 | Paragraph 5 of management strategy | 0.25 | 0.24 | 307208 | 12 | 15 |
| Precautionary approach | 224186 | 212434 | 192465 | 19969 | $\begin{aligned} & \text { SSB }=\min \{1 ; \\ & \text { SSB2017/Btrigger }\} \end{aligned}$ | 0.87 | 0.83 | 169000 | -38 | 210 |
| Zero catch | 0 | 0 | 0 | 0 | $\mathrm{F}=0$ | 0 | 0 | 391963 | 43 | -100 |
| Other options | 79803 | 76084 | 68932 | 7152 | F2016 | 0.24 | 0.23 | 310253 | 13 | 11 |
|  | 72198 | 68601 | 62153 | 6448 | TAC2016 | 0.21 | 0.20 | 315995 | 15 | 0 |
|  | 115286 | 109922 | 99589 | 10333 | Fpa | 0.37 | 0.35 | 273885 | 0 | 60 |

Table A9. Saithe in Subareas 4 and 6 and Division 3a. The basis for the catch options.

| VARIABLE | VALUE | SourCe | Notes |
| :--- | :--- | :--- | :--- |
| F ages 4-7 (2016) | 68601 t | ICES <br> $(2016 a)$ | TAC constraint (F=0.24) |
| SSB (2016) | 242142 t | ICES <br> $(2016 a)$ | SSB in the intermediate year |
| SSB (2017) | 274310 t | ICES <br> $(2016 a)$ | SSB at the beginning of the TAC year |
| Rage3 (2016) | 99 billion | ICES <br> $(2016 a)$ | Median recruitment resampled from 2003- <br> 2015 |
| Rage3 (2017) | 99 billion | ICES <br> $(2016 a)$ | Median recruitment resampled from 2003- <br> 2015 |
| Total catch (2016) | 72064 t | ICES <br> $(2016 a)$ | Assuming 2015 landings fraction by age |
| Commercial | 68601 t | ICES <br> $(2016 a)$ | TAC 2015 |
| landings (2016) | 3463 t | ICES <br> $(2016 a)$ | Assuming 2015 discard fraction by age |
| Discards (2016) |  |  |  |

# Annex 10 Technical Minutes of the Review Group of Precautionary Approach Reference Points estimation 

Review of ICES WGNSSK Report 2016
9 May 2016-31 May 2016
Reviewers: Chris Legault (chair)
Arni Magnusson
Colin Millar
Chair WG: Alexander Kempf (Germany) and José De Oliveira (UK)

Secretariat: Cristina Morgado

## General

The RG acknowledges the intense effort expended by the working group to produce the report.
The Review Group considered estimation of PA reference points for the following stocks:
Cod in Subarea 4 and Divisions 7.d and 3.a (North Sea. Eastern English Channel. Skagerrak)

Haddock in Subarea 4 and Divisions 3.a and 6.a (North Sea. Skagerrak and West of Scotland)

Plaice in Division 7.d (Eastern Channel)
Sole in Subarea 4 (North Sea)
Whiting Subarea 4 (North Sea) and Division 7.d (Eastern Channel)
Plaice in Subarea 4(North Sea) and Division 3.a. 20 (Skagerrak)

## Cod in Subarea 4 and Divisions 7.d and 3.a (report section 14.9)

## General comments

According to the advice sheet, $\mathrm{B}_{\lim }=118 \mathrm{kt}$ and $\mathrm{B}_{\mathrm{pa}}=165 \mathrm{kt}$, and the basis of $\mathrm{B}_{\mathrm{lim}}$ is the SSB associated with the last above-average recruitment (1996 year class). The use of 1.4 assumes $\sigma_{B}$ $=0.20$.

According to the advice sheet, $\mathrm{F}_{\mathrm{lim}}=0.58$ and $\mathrm{F}_{\mathrm{pa}}=0.41$. The basis of $\mathrm{F}_{\mathrm{lim}}$ is an EQSim analysis based on recruitment period 1998-2014, where the changepoint of the segmented regression was estimated rather than forced at $\mathrm{B}_{\text {lim, }}$, this deviation from the ices guidelines is justified in the report. The use of 1.4 assumes $\sigma_{\mathrm{F}}=0.20$. Estimation uncertainty in final year F gives $\sigma_{\mathrm{F}}=0.12$, but the WG considered this too low.

The value 1.4 has historically been used to derive PA reference points, the underlying logic is an assumption of $\sigma=0.20$. For consistency with the guidelines on PA reference points, the value of $\sigma_{B}$ and $\sigma_{F}$ should be made explicit in the advice sheet along with the equation $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\mathrm{lim}} * \exp \left(1.645 \sigma_{\mathrm{B}}\right)$ or $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} * \exp \left(-1.645 \sigma_{\mathrm{F}}\right)$.

The argument used to define $\mathrm{B}_{\text {lim }}$ is sound, but it was difficult to review without a stock-recruitment plot.

It was difficult to review the argument used to define $\mathrm{F}_{\text {lim }}$ without a stock-recruitment plot.

## Technical comments

|  | Basis of underlying limit refpt is clear | Right approach to derive PA refpt from limit refpt | PA refpt looks correct | Basis and value of $\sigma$ is clear |
| :---: | :---: | :---: | :---: | :---: |
| $\mathrm{B}_{\mathrm{pa}}$ | OK, (type 1 applied to recent time series) | OK | OK | No, value of $\sigma_{B}$ not stated. <br> Please state the formula <br> $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\mathrm{lim}} * \exp (1$. $645 \sigma$ ) and state the value of $\sigma$ used. |
| $\mathrm{F}_{\mathrm{pa}}$ | OK, method a) stochastic simulation was used to estimate $\mathrm{F}_{\text {lim }}$. | OK | OK | No, value of $\sigma_{\mathrm{F}}$ not stated. <br> Please state the formula $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\mathrm{lim}} * \exp (-$ <br> $1.645 \sigma$ ) and state the value of $\sigma$ used. |

## Conclusions

$2 / 8$ cells require attention. Small revisions to the advice sheet are required to address this.

## Haddock in Subarea 4 and Divisions 3.a and 6.a (report section 13)

## General comments

According to the advice sheet, $\mathrm{B}_{\mathrm{lim}}=96$ and $\mathrm{B}_{\mathrm{pa}}=134$. The basis of $\mathrm{B}_{\mathrm{lim}}$ is the lowest estimated SSB which resulted in high recruitment (1972), and the basis of $\mathrm{B}_{\mathrm{pa}}$ is $1.4 * \mathrm{~B}_{\mathrm{lim}}$.

According to the advice sheet, $\mathrm{F}_{\mathrm{lim}}=0.384$ and $\mathrm{F}_{\mathrm{pa}}=0.274$. The basis of $\mathrm{F}_{\text {lim }}$ is estimation by application of EqSIM evaluation, and the basis of $\mathrm{F}_{\mathrm{pa}}$ is stated in the advice sheet to be also estimation by application of EqSIM evaluation. From the values, however, it seems that the basis of $\mathrm{F}_{\mathrm{pa}}$ is $\mathrm{F}_{\text {lim }} / 1.4$, and this is the basis stated in the report.

## Technical comments

|  | Basis of underlying limit refpt is clear | Right approach to derive PA refpt from limit refpt | PA refpt looks correct | Basis and value of $\sigma$ is clear |
| :---: | :---: | :---: | :---: | :---: |
| $\mathrm{B}_{\mathrm{pa}}$ | $\mathrm{OK}, \mathrm{B}_{\lim }$ is the lowest biomass where large recruitment is observed (type 1) | Please state the equation $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\lim } * \exp (1.64$ <br> $5 \sigma_{\mathrm{B}}$ ) and the value of $\sigma_{B}=0.20$ in the refpt table of the advice sheet. | Almost, but $\mathrm{B}_{\lim } * \exp \left(1.645^{*} 0\right.$. 20) will give a slightly different value. | Please state the value of $\sigma_{B}=0.20$ in the refpt table of the advice sheet. |
| $\mathrm{F}_{\mathrm{pa}}$ | OK, method A (stochastic simulations) was used to evaluate $\mathrm{F}_{\text {lim }}$. | Please state the equation $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\mathrm{lim}} * \exp (-$ <br> $1.645 \sigma_{\mathrm{F}}$ ) and the value of $\sigma_{F}$ in the refpt table of the advice sheet. | Almost, but $\mathrm{F}_{\text {lim }}$ *exp(- <br> $1.645 * 0.20$ ) will give a slightly different value. | Please state the value of $\sigma_{F}=0.20$ in the refpt table of the advice sheet. |

## Conclusions

$2 / 8$ cells in the above matrix are OK. The remaining 6 should be improved.

## Plaice in Division 7.d (report section 6)

## General comments

According to the advice sheet, $\mathrm{B}_{\mathrm{lim}}=18448$ based on the break point of the segmented regression SRR and $\mathrm{B}_{\mathrm{pa}}=25826$ based on the relationship $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\lim } * 1.4$.

According to the advice sheet, $\mathrm{F}_{\text {lim }}=0.5$ based on EqSim that will maintain the stock above $\mathrm{B}_{\text {lim }}$ with a $50 \%$ probability and $\mathrm{F}_{\mathrm{pa}}=0.36$ based on the relationship $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} / 1.4$.

## Technical comments

|  | Basis of underlying limit refpt is clear | Right approach to derive PA refpt from limit refpt | PA refpt looks correct | Basis and value of $\sigma$ is clear |
| :---: | :---: | :---: | :---: | :---: |
| $\mathrm{B}_{\mathrm{pa}}$ | $\mathrm{OK}, \mathrm{B}_{\mathrm{lim}}$ is estimated from segmented regression (Type 2) | Please state the equation $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\lim } * \exp (1.64$ <br> $5 \sigma_{\mathrm{B}}$ ) and the value of $\sigma_{B}=0.20$ in the refpt table of the advice sheet. | Almost, but $\mathrm{B}_{\mathrm{lim}} * \exp (1.645 * 0$. 20) will give a slightly different value. | Please state the value of $\sigma_{B}=0.20$ in the refpt table of the advice sheet. |
| $\mathrm{F}_{\mathrm{pa}}$ | OK, method A (stochastic simulations) was used to evaluate $\mathrm{F}_{\text {lim }}$. | Please state the equation $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\mathrm{lim}}{ }^{*} \exp (-$ $1.645 \sigma_{\mathrm{F}}$ ) and the value of $\sigma_{\mathrm{F}}=0.20$ in the refpt table of the advice sheet. | OK | Please state the value of $\sigma_{F}=0.20$ in the refpt table of the advice sheet. |

## Conclusions

$3 / 8$ cells in the above matrix are OK. The remaining 5 should be improved.

## Sole in Subarea 4 (report section 10.7)

## General comments

According to the advice sheet, $\mathrm{B}_{\lim }=26.3$ and $\mathrm{B}_{\mathrm{pa}}=37.0$. The basis of $\mathrm{B}_{\text {lim }}$ is $\mathrm{B}_{\text {loss }}$ and the basis of $\mathrm{B}_{\mathrm{pa}}$ is $1.4 * \mathrm{~B}_{\mathrm{lim}}$. This implies $\sigma_{\mathrm{B}}=0.20$ and $\mathrm{B}_{\mathrm{pa}}$ has been rounded from 36.8 to 37.0 .

According to the advice sheet, $\mathrm{F}_{\text {lim }}=0.63$ and $\mathrm{F}_{\mathrm{pa}}=0.44$. The basis of $\mathrm{F}_{\text {lim }}$ is the F that in equilibrium will maintain the stock above $\mathrm{B}_{\lim }$ with a $50 \%$ probability, and the basis of $\mathrm{F}_{\mathrm{pa}}$ is $\mathrm{F}_{\mathrm{lim}} / 1.4$. This implies that $\sigma_{\mathrm{F}}=0.20$.

A closer examination of the $\mathrm{F}_{\text {lim }}$ analysis (Sharepoint file WGNSSK/Reference points Flim and Fpa/sol-nsea/Ref_points.7z) reveals that the $\mathrm{F}_{\text {lim }}$ was calculated as 0.623 , which could be rounded to 0.62 , but not 0.63 .

## Technical comments

|  | Basis of underlying limit refpt is clear | Right approach to derive PA refpt from limit refpt | PA refpt looks correct | Basis and value of $\sigma$ is clear |
| :---: | :---: | :---: | :---: | :---: |
| $\mathrm{B}_{\mathrm{pa}}$ | $\begin{aligned} & \text { OK, } \quad B_{\text {lim }}=B_{\text {loss }} \\ & \text { (type 5) } \end{aligned}$ | Please state the equation $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\mathrm{lim}} * \exp (1.64$ $5 \sigma_{\mathrm{B}}$ ) and the value of $\sigma_{B}=0.20$ in the refpt table of the advice sheet. | OK, if 0.20 is the best estimate of $\sigma_{B}$. | Please state the value of $\sigma_{\mathrm{B}}=0.20$ in the refpt table of the advice sheet. |
| $\mathrm{F}_{\mathrm{pa}}$ | Method A (stochastic simulations) was used to evaluate $\quad \mathrm{F}_{\text {lim }}$. However, the resulting $F_{\text {lim }}$ was probably 0.62 , not 0.63. | Please state the equation $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\mathrm{lim}}{ }^{*} \exp (-$ $1.645 \sigma_{\mathrm{F}}$ ) and the value of $\sigma_{F}$ in the refpt table of the advice sheet. | No, if $\mathrm{F}_{\text {lim }}$ is 0.62 (or 0.63 , for that matter) then $\mathrm{F}_{\mathrm{pa}}$ should be 0.45 . | Please state the value of $\sigma_{\mathrm{F}}=0.20$ in the refpt table of the advice sheet. |

## Conclusions

$2 / 8$ cells in the above matrix are OK . The remaining 6 should be improved.

## Whiting Subarea 4 and Division 7.d (report section 12.1.8)

## General comments

According to the advice sheet, $\mathrm{B}_{\text {lim }}=173 \mathrm{kt}$ and $\mathrm{B}_{\mathrm{pa}}=242 \mathrm{kt}$. The basis of $\mathrm{B}_{\text {lim }}$ is $\mathrm{B}_{\text {loss }}$ (SSB in 2007 in the 2016 assessment) and the basis of $\mathrm{B}_{\mathrm{pa}}$ is $1.4 * \mathrm{~B}_{\text {lim }}$. This implies $\sigma_{\mathrm{B}}=0.20$.

According to the advice sheet, $\mathrm{F}_{\mathrm{lim}}=0.39$ and $\mathrm{F}_{\mathrm{pa}}=0.28$. The basis of $\mathrm{F}_{\text {lim }}$ is Eqsim $\left(\mathrm{F}_{50}\right)$, and the basis of $\mathrm{F}_{\mathrm{pa}}$ is $\mathrm{F}_{\mathrm{lim}} / 1.4$. This implies that $\sigma_{\mathrm{F}}=0.20$.

When using the value 1.4 to derive PA reference points, the underlying logic is an assumption of $\sigma=0.20$. For consistency with the guidelines on PA reference points, the value of $\sigma_{B}$ and $\sigma_{F}$ should be made explicit in the advice sheet along with the equation $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\lim } * \exp \left(1.645 \sigma_{\mathrm{B}}\right)$ or $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} * \exp \left(-1.645 \sigma_{\mathrm{F}}\right)$.

## Technical comments

|  | Basis of underlying limit refpt is clear | Right approach to derive PA refpt from limit refpt | PA refpt looks correct | Basis and value of $\sigma$ is clear |
| :---: | :---: | :---: | :---: | :---: |
| $\mathrm{B}_{\mathrm{pa}}$ | $\begin{aligned} & \text { OK, } \quad B_{\text {lim }}=B_{\text {loss }} \\ & \text { (type 5) } \end{aligned}$ | Please state the equation $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\lim } * \exp (1.64$ <br> $5 \sigma_{\mathrm{B}}$ ) and the value of $\sigma_{B}=0.20$ in the refpt table of the advice sheet. | Almost, but $\mathrm{B}_{\mathrm{lim}} * \exp (1.645 * 0$. 20) will give a slightly different value (at $3^{\text {rd }}$ sig. fig.). | Please state the value of $\sigma_{B}=0.20$ in the refpt table of the advice sheet. |
| $\mathrm{F}_{\mathrm{pa}}$ | OK, method a) stochastic simulation was used to estimate $\mathrm{F}_{\text {lim }}$. | Please state the equation $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\mathrm{lim}} * \exp (-$ <br> $1.645 \sigma_{\mathrm{F}}$ ) and the value of $\sigma_{F}$ in the refpt table of the advice sheet. | OK | Please state the value of $\sigma_{F}=0.20$ in the refpt table of the advice sheet. |

## Conclusions

$3 / 8$ cells in the above matrix are OK. The remaining 5 should be improved.

## Plaice in Subarea 4 and Division 3.a. 20 (report section 8.6)

## General comments

According to the advice sheet, $\mathrm{B}_{\text {lim }}=160 \mathrm{kt}$ and $\mathrm{B}_{\mathrm{pa}}=230 \mathrm{kt}$. The basis of $\mathrm{B}_{\text {lim }}$ is $\mathrm{B}_{\text {loss }}$, the lowest observed biomass in 1997 as assessed in 2004, and the basis of Bpa is $1.44 \times \mathrm{B}_{\text {lim.. }}$.

According to the advice sheet, $\mathrm{F}_{\mathrm{lim}}=0.63$ and $\mathrm{F}_{\mathrm{pa}}=0.45$. The basis of Fpa is the F that in equilibrium will maintain the stock above $\mathrm{B}_{\lim }$ with a $50 \%$ probability, and the basis of $\mathrm{F}_{\mathrm{pa}}$ is $\mathrm{F}_{\text {lim }} / 1.4$.

The value 1.4 is often used to derive PA reference points, the underlying logic is an assumption of $\sigma=0.20$. For consistency with the guidelines on PA reference points, the value of $\sigma_{B}$ and $\sigma_{F}$ should be made explicit in the advice sheet along with the equation $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\lim } * \exp \left(1.645 \sigma_{\mathrm{B}}\right)$ or $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} * \exp \left(-1.645 \sigma_{\mathrm{F}}\right)$.

## Technical comments

|  | Basis of underlying limit refpt is clear | Right approach to derive PA refpt from limit refpt | PA refpt looks correct | Basis and value of $\sigma$ is clear |
| :---: | :---: | :---: | :---: | :---: |
| $\mathrm{B}_{\mathrm{pa}}$ | OK, $\quad B_{\text {lim }}=B_{\text {loss }}$ (type 5) | Please state the equation $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\mathrm{lim}} * \exp (1.64$ <br> $5 \sigma_{\mathrm{B}}$ ) and the value of $\sigma_{B}=0.22$ in the refpt table of the advice sheet. | OK | Please state the value of $\sigma_{B}$ in the refpt table of the advice sheet. 1.44 implies $\sigma_{B}=0.22$ |
| $\mathrm{F}_{\mathrm{pa}}$ | OK, method a) stochastic simulation was used to estimate $\mathrm{F}_{\text {lim }}$. | Please state the equation $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\mathrm{lim}} * \exp (-$ <br> $1.645 \sigma_{\mathrm{F}}$ ) and the value of $\sigma_{\mathrm{F}}=0.20$ in the refpt table of the advice sheet. | OK | Please state the value of $\sigma_{F}=0.20$ in the refpt table of the advice sheet. |

## Conclusions

4 out of 8 cells are OK, the remaining 4 should be addressed.

## Saithe in Subareas 4 and 6 and Division 3.a (report section 11)

## General comments

According to the advice sheet, $\mathrm{B}_{\lim }=107$ and $\mathrm{B}_{\mathrm{pa}}=150$. The basis of $\mathrm{B}_{\text {lim }}$ is $\mathrm{B}_{\text {loss }}$ and the basis of $\mathrm{B}_{\mathrm{pa}}$ is $\mathrm{B}_{\text {lim }} * \exp \left(1.645 \sigma_{\mathrm{B}}\right)$, where $\sigma_{\mathrm{B}}=0.20$.

According to the advice sheet, $\mathrm{F}_{\text {lim }}=0.56$ and $\mathrm{F}_{\mathrm{pa}}=0.40$. The basis of $\mathrm{F}_{\text {lim }}$ is the F that gives $50 \%$ probability to fall below Blim in the stochastic EqSim simulations, and the basis of $\mathrm{F}_{\mathrm{pa}}$ is $\mathrm{F}_{\text {lim }} * \exp \left(-1.645 \sigma_{\mathrm{F}}\right)$, where $\sigma_{\mathrm{F}}=0.20$.

## Technical comments

|  | Basis of underlying limit refpt is clear |  | PA refpt looks correct | Basis and value of $\sigma$ is clear |
| :---: | :---: | :---: | :---: | :---: |
| $\mathrm{B}_{\mathrm{pa}}$ | $\begin{aligned} & \text { OK, } \\ & \text { (type 5) } \end{aligned} \quad B_{\text {lim }}=B_{\text {loss }}$ | OK | OK | OK |
| $\mathrm{F}_{\mathrm{pa}}$ | OK, method A (stochastic simulations) was used to evaluate $\mathrm{F}_{\text {lim }}$. | OK | OK | OK |

## Conclusions

$8 / 8$ cells in the above matrix are OK. The PA reference points have been evaluated according to the guidelines.


[^0]:    * Wanted" and "unwanted" catch are used to described Nephrops that would be landed and discarded in the absence of the EU landing obligation based on discard rates estimates for average (2013-2015).
    ** Calculated for dead removals and applied to total catch.

[^1]:    * provisional na = not available
    ** Totals for 1991-94 exclusive of landings by the Netherlands

[^2]:    * provisional na $=$ not available
    ** There are no landings by other countries from this FU

[^3]:    * provisional na = not available
    ** Totals for 1993-94 exclusive of landings by the Netherlands

[^4]:    * provisional

[^5]:    * "Wanted" and "unwanted" catch are used to describe fish that would be landed and discarded in the absence of the EU landing obligation, based on discard rates estimates for 2015.

[^6]:    *Preliminary

[^7]:    * Preliminary

[^8]:    ${ }^{1)}$ No data reported by France in 1999.

[^9]:    * "Wanted" and "unwanted" catch are used to describe fish that would be landed and discarded in the absence of the EU landing obligation, based on discard rates estimates for 2015.
    ** Wanted catch split according to the average in 1993-1998, i.e. 90.6\% in Subarea 4 and Subdivision 3.a.20 and 9.4\% in Subarea 6.
    *** SSB 2018 relative to SSB 2017

[^10]:    * Assumed for all fleets
    ** Unwanted landings are those animals >MCS but historically discarded
    *** Calculated for dead removals

[^11]:    $N_{s}$ mean abundance in the strata $s$, expressed in number/km²

